High Resolution Water Table Modeling of the Shallow Groundwater Using a Knowledge-Guided Gradient Boosting Decision Tree Model

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Detailed knowledge of the uppermost water table representing the shallow groundwater system is critical in order to address societal challenges that relate to the mitigation and adaptation to climate change and enhancing climate resilience in general. Machine learning (ML) allows for high resolution modeling of the water table depth beyond the capabilities of conventional numerical physically-based hydrological models with respect to spatial resolution and overall accuracy. For this, in-situ well and proxy observations are used as training data in combination with high resolution covariates. The objective of this study is to model the depth of the uppermost water table for a typical summer and winter condition at 10 m spatial resolution over entire Denmark (43,000 km²). CatBoost, a state of the art implementation of gradient boosting decision trees, is employed in this study to model the water table depth and the associated uncertainties. The groundwater domain has not been the most prominent field of applications of recent hydrological ML advances due to the lack of big data. This study brings forward a novel knowledge-guided ML framework to overcome this limitation by integrating simulation results from a physically-based groundwater flow model. The simulation data are utilized to (1) identify wells that represent the uppermost water table, (2) augment missing training data by accounting for simulated water level seasonality, and (3) expand the list of covariates. The curated training dataset contains around 13,000 wells, 19,000 groundwater proxy observations at lakes, streams and coastline as well as 15 covariates. Cross validation attests that the ML model generalizes well with a mean absolute error of around 115 cm considering solely well observations and a MAE of <50 cm taking also the proxy observations into consideration. Quantile regression is applied to estimate confidence intervals and the estimated uncertainty is largest for moraine clay soils that are characterized with a distinct geological heterogeneity. This study highlights a novel research avenue of knowledge-guided ML for the groundwater domain by efficiently supporting a ML model with a physically-based hydrological model to predict the depth of the water table at unprecedented spatial detail and accuracy.

Keywords: water table depth, machine learning, high resolution, CatBoost, quantile regression
INTRODUCTION

A key state variable of the hydrological cycle is the depth of the uppermost water table, i.e., shallow groundwater, with a broad range of crucial societal, and environmental implications such as securing infrastructure, food production and sustaining ecosystems (Gleeson et al., 2016). Following the global analysis of water table patterns by Fan et al. (2013), up to one-third of the land area is influenced by shallow groundwater, being either directly groundwater-fed or having the water table or capillary fringe within plant rooting depths. More concretely, the shallow groundwater system plays a key role in building mitigation and adaptation measures to address climate change, as the uppermost water table controls greenhouse gas emissions from wetlands (Tiemeyer et al., 2016, 2020) and as rising groundwater exacerbates flooding and falling groundwater intensifies droughts (Taylor et al., 2013). Floods can either be directly induced or intensified by the shallow groundwater system which has a special relevance for urban areas (MacDonald et al., 2012; Bricker et al., 2017). In an agronomic context, the shallow groundwater system is vital to meet crop water requirements in many agricultural settings while having adverse consequences for crop yield when the water table is too close to the surface (Kahlow et al., 2005; Zipper et al., 2015). Moreover, the uppermost water table constitutes a link between subsurface and the land-surface by affecting the energy balance and near-surface climatic conditions (Larsen et al., 2016; Maxwell and Condon, 2016).

The above mentioned relevancy of the shallow groundwater system requires versatile modeling systems to meet the demands of decision makers with respect to accuracy and spatial resolution. Accuracy and spatial resolution can be considered main bottlenecks in the advancement of physically-based numerical hydrological models. Increasing the spatial resolution does not necessarily go along with an improvement of the model accuracy, as established process descriptions and parametrizations are not always scalable (Beven and Cloke, 2012; Clark et al., 2017). At the same time, increasing the number of computational units results in a computational burden limiting the possibility to conduct thorough parameter calibration and sensitivity analysis. Due to the stringent parametrization and rigid model structure, conventional hydrological models cannot fully harness the wealth of readily available environmental big data. Nevertheless, physically-based models integrate decades of hydrological knowledge and are indispensable for integrated assessments of the hydrological cycle and simulating hydrological response under non-stationarity, i.e., climate change. As outlined by Shen (2018), Reichstein et al. (2019), and Nearing et al. (2020), machine learning (ML) is gaining increasing attention in the hydrological science to overcome some of the previously mentioned constraints of conventional physically-based models. ML has the advantage of being data flexible with respect to optimally utilizing the wealth of environmental big data while providing accurate predictions at low computational costs. However, ML lacks process descriptions which commonly restricts trained ML models to deliver predictions within observed ranges of the training dataset. In order to reconcile advantages of both modeling perceptions, knowledge-guided ML forms a promising new research avenue (Rajaei et al., 2019; Kraft et al., 2020). Knowledge-guided ML has the aim to integrate physical consistency into ML improve model performance and robustness. There exist multiple approaches to design knowledge-guided ML models as outlined by Read et al. (2019), Reichstein et al. (2019), Konapala et al. (2020), and others, which are also referred to as physics- or process-guided. A clear formal definition of these modeling approaches is still lacking, but they generally aim at integrating aspects of scientific knowledge into a ML model.

This study aims at modeling the depth of the uppermost water table at 10 m spatial resolution over entire Denmark for a typical summer and winter condition by implementing a knowledge-guided ML model. The physically-based information are obtained from the Danish national water resources model, that integrates groundwater and surface water processes (Højberg et al., 2013; Stisen et al., 2019). The physically-based model is used three-fold, (1) to derive threshold depths to select wells that reflect the shallow groundwater, (2) to augment training data at wells with incomplete pairs of summer-winter observations, and (3) to inform the ML model with the typical summer and winter water table depth using simulation results at 100 m resolution.

Decision tree based ML models are popular tools for geospatial modeling of environmental variables (Hengl et al., 2018; Tyralis et al., 2019b). In this context a target variable, available as point data, is used in conjunction with maps of explanatory variables to curate a training dataset. Relationships between the target variable and the explanatory variables are established via the decision trees which, once trained, can be generalized to make predictions of the target variable at all grids. This framework has been successfully applied across the geosciences to model water chemistry indicators (Tesoriero et al., 2015; Erickson et al., 2021) soil properties (Moller et al., 2017; Hengl et al., 2021), subsurface redox conditions (Close et al., 2016; Koch et al., 2019a), water table depth (Bachtold et al., 2014; Koch et al., 2019b), and other variables. In such modeling frameworks, uncertainty can be quantified via quantile regression (López López et al., 2014; Tyralis et al., 2019a). Gradient boosting is among the state of the art techniques to build decision tree models and, for this study we have employed the CatBoost implementation of gradient boosting decision trees (Dorogush et al., 2018; Prokhorenkova et al., 2018). Numerous studies apply ML to model the temporal water table dynamics by using various ML techniques (Sun et al., 2016; Guzman et al., 2017; Wunsch et al., 2021). These studies highlight successful applications, but are always limited to time series modeling, at a few selected sites. Despite these efforts, the spatial dimension is often neglected in the published studies. Fienen et al. (2013), Bechtold et al. (2014), and Koch et al. (2019b) are among the few studies that model the spatial variability of water table depth using ML. Knowledge-guided ML applications in the groundwater domain are for example the physics informed neural networks model developed by Guo et al. (2020) that allows to solve partial differential equations with less calculational time.

The three main objectives of the paper are as follows: (1) to train a ML model to predict the uppermost water table depth
at 10 m spatial resolution over entire Denmark, (2) to quantify uncertainty using quantile regression, and (3) to formalize a knowledge-guided ML framework that builds upon a physically-based hydrological model.

**MATERIALS AND METHODS**

**Study Area**

This study is carried out for the entire land phase of Denmark, located in Northern Europe and covering an area of around 43,000 km$^2$ (Figure 1). Denmark is generally flat with a maximum elevation of 170 m.a.s.l. and agriculture is the main land cover with around 70%. The landscape of Denmark was formed by a sequence of Pleistocene glaciations and postglacial processes. The soils of eastern Denmark are dominated by Weichselian moraine sediments with a moderate clay content, whereas western Denmark is characterized by older moraine sediments originating from the Saalian age intertwined by sandy Weichselian outwash plains.

In Denmark, the groundwater system is under pressure as a consequence of climate change and abstractions, revealed by quantitative modeling assessments conducted by Henriksen et al. (2008) and Karlsson et al. (2016). More specifically, a shallow water table rise of up to 1.5 m for a 100-year event relative to present average conditions was estimated by Kidmose et al. (2013) for an urban catchment.

**Data**

In order to make seasonal estimates (typical winter and typical summer) of the shallow water table at high resolution, an initial comprehensive data processing has been conducted to curate a high-quality dataset containing water level observations at shallow wells, additional groundwater observations as well as national maps of explanatory variables.

**Physically-Based Model**

The national water resources model of Denmark (DK-model), which has previously been refined from the original 500 m resolution to an updated 100 m resolution, was employed in this study (Højberg et al., 2013; Stisen et al., 2019; Henriksen et al., 2020). The core of the DK-model is an integrated surface-subsurface hydrological model (MIKE SHE model code) that is calibrated against streamflow and groundwater observations. We utilized simulation results of the depth of the uppermost water table at daily time step from a 30 year simulation period (1990–2019) and processed the data to the maps depicted in Figure 2.

The summer condition reflects a median across 30 summers, each summer defined as the median of the months June, July, and August. Analogous, the winter condition is based on the median
of 30 winters, where each winter is defined as the median of the months December, January, and February of the same winter season. The third processed variable is the simulated median seasonality, which is based on the median across 30 annual seasonalities. For each year, the seasonality has been calculated as median summer condition minus median winter condition. The simulated summer condition has a mean water table depth of 3.5 m whereas the simulated winter condition has a mean of 2.6 m, with a standard deviation of 5.8 and 4.8 m, respectively.

In terms of the spatial pattern, the simulated water table depth shows overall resemblance between the summer and the winter condition. Deepest groundwater levels are found in the Western part of Denmark alongside high elevation and for sandy soils. The moraine clay soils are characterized with a dominant shallow water table, especially during the winter months. The simulated median seasonality has a mean of 85 cm. The spatial pattern of the simulated seasonality reflects a complex interplay of geology and topography. The amplitude is generally low in topographical sinks and high for areas with an elevated topography. A majority of the moraine clay soils possess an amplitude larger than 1 m, which is caused by a winter water table at the surface due to slow infiltration in combination with a drying out in the summer months.

**Groundwater Observations**

The open-access Danish well database (Jupiter) contains around 100,000 wells with at least a single water table observation in the selected 30-year period between 1990 and 2019. Based on the national well dataset, we first identified the wells that represent the shallow groundwater system, i.e., characterize the uppermost water table. Given the geological complexity of Denmark, this can vary from just a few meters for locations with a thick surficial clay layer to several tens of meters for the sandy outwash plains.

In order to find suitable threshold depths, that classify a well as being either shallow or not, we combined the national soil map (Figure 1) and the simulated uppermost water table (Figure 2). For this analysis, the midpoint intake depth of each well was used, which is relative to the top and the bottom of the intake. The 95th percentile of the simulated water table depth was calculated for each soil type which guided the definition of the presented threshold depths applied to the midpoint intake depths of the wells (Table 1). Wells were only selected if the intake depth was lower than the soil type dependent threshold depth. A spatially soil type distributed threshold depth, guided by a physically-based model, has the advantage to reflect the natural conditions best possibly. Alternative, a constant threshold depth of e.g., 10 m would result in the selection of wells that do not reflect the uppermost water table in clayey settings where a well may be placed in a sand unit below a 6 m surficial clay layer containing the uppermost water table.

After applying the soil types dependent threshold depths, summer, and winter depths to the uppermost groundwater were calculated at each well. For this task, summer refers to observations from the months June, July, and August (JJA) whereas winter refers to the months December, January, and February (DJF). Inter-annual variation was not considered during this processing step and in case a well contained several summer or winter observations a median was calculated. This resulted in 13,047 well, as shown in Figure 3, of which 1,378 wells had both a summer and a winter observation, 5,651 only a winter observation and 6,018 only a summer observation. For the wells that were missing either a winter or a summer observation the simulated median seasonality (Figure 2) was used to augment the missing season, i.e., winter...
= summer—seasonality, and vice versa. Further, the training dataset was extended by additional observations that reflect proxy observations for surficial groundwater levels with a depth of zero. In total, 19,074 groundwater connected lakes with an aerial extent of at least 100 m² were used for this purpose. Additionally, 1,000 points, randomly placed along each, the stream network and along the coastline, were added to better represent these under sampled settings where the water table is expected to be at the surface year-round. This resulted in a total number of 34,061 groundwater observations spread across Denmark, yielding a density of around 0.8 observations per km². Based on the well data only, the average summer depth is 3.6 and 3.1 m for the winter season with a standard deviation of 2.8 m for both.

**Explanatory Variables**

In Table 2, an overview of the covariates used to model the depth of the uppermost water table is presented. In total, 15 covariates were assembled as input to the ML model. This list comprises information on soil texture, geology, topography-based characteristics, water body proximity, land cover, and outputs from a hydrological simulation with the DK-model. The native spatial resolution of the covariates varied, but all covariates were resampled to 10 m to be in agreement with the defined output resolution. For the resampling we used a bilinear interpolation method for the continuous variables. The water body proximity was expressed as both the vertical and horizontal distance to the nearest water body, which contained rivers, lakes, and the coastline.

**Gradient Boosting Decision Trees**

We applied a new implementation of the gradient boosting decision tree (GBDT) algorithm, i.e., Cat Boost that was first developed by Yandex engineers in 2017 (Dorogush et al., 2018; Prokorenkova et al., 2018). GBDT, first introduced by Friedman (2001), is a popular ML method that is increasingly being applied across the geosciences for both, classification, and regression tasks (Fan et al., 2018; Georganos et al., 2018; Moller et al., 2018). GBDT builds a prediction model based on an ensemble of weak learners, i.e., decision trees. In an additive training process, GBDT attempts to correct itself by adding a decision tree trained against the residuals of the ensemble sum of its predecessors for a pre-defined number of iterations. Small incremental steps are taken to correct the residuals, which is controlled via the learning rate. The learning rate is a key hyper parameter of GBDT that is multiplied with the predicted residuals of each decision tree. In order to further alleviate overfitting, stochasticity is added to the training process via subsampling of the training dataset each time a new tree is built. Also, CatBoost allows for early stopping of the training process once the objective function for a test dataset stagnates for a defined number of iterations. The individual decision trees are constrained by a number of hyper parameters, such as tree depth or the minimum number observations per split. CatBoost is specifically designed to work well with categorical variables and a preprocessing of categorical variables by means of different encoding methods is not necessary. Also, CatBoost is favorable over similar ML algorithms, such as Random Forests, Support Vector Machines, or other GBDT implementations (e.g., XGBoost or LightGBM), with respect to computational time and memory usage, while achieving a competitive accuracy (Huang et al., 2019; Hancock and Khoshgoftaar, 2020). Given the size of the training dataset (34,061 grids and 15 variables) and the prediction dataset (430 million grids and 15 variables) the memory efficiency and the overall computational time were crucial parameters for the selection of the ML algorithm. The covariate importance of a trained CatBoost model can be revealed through analysis of the tree splits to calculate how much the prediction changes on average if the feature value changes.

For the purpose of simulating the depth of the uppermost water table using CatBoost we employ two different objective functions. First, the mean absolute error (MAE) is used to train the best estimate of the winter and summer condition. One model is trained for each season. The MAE is expressed as follows:

$$\text{MAE} = \frac{\sum_{i=1}^{n} |\text{sim}_i - \text{obs}_i|}{n}$$

where sim is the simulated groundwater depth and obs the observed for a total of n training data. Besides simulating the best estimate, we are also interested in quantifying the uncertainty of the model. For this, we utilized quantile regression to define objective functions targeting specific quantiles of the distribution. This yields a probabilistic model that is not trained to estimate the conditional mean, but to estimate a defined quantile q of the distribution instead:

$$\text{Quantile} = \frac{\sum_{i=1}^{n} \left( q - 1 \cdot (\text{obs}_i \leq \text{sim}_i) \right) \cdot (\text{obs}_i - \text{sim}_i)}{n}$$

Setting q to 0.5 yields the same result as the MAE, but setting q to other values will give asymmetric weights to the residual depending on q and the overall sign of the error. Setting q to 0.1 will estimate the 10th percentile, by associating a weight of 0.9 to over predictions and a weight of 0.1 to under predictions. Thereby the 10th percentile can be approximated, meaning that the model will be trained to over predict 90% of the times. With the uncertainty analysis we intend to estimate the 68 and

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### TABLE 1 | Threshold depths that define a shallow well with respect to soil types (Figure 1).

| Soil Type       | Max depth (m.b.g.l.) |
|-----------------|----------------------|
| Sand            | 15                   |
| Moraine Sand    | 20                   |
| Peat            | 3                    |
| Chalk           | 20                   |
| Unclassified    | 3                    |
| Moraine Clay    | 3                    |
| Sand and Gravel | 10                   |
| Marine Sand     | 3                    |
| Silt            | 3                    |

Values of max depth were derived from the 95th percentile of the simulated water table depths for each of the given soil types and were applied to the midpoint intake depths of the wells.
FIGURE 3 | Training dataset containing wells (left panel) and additional observations (right panel). The dataset contains 13,047 shallow wells with both a winter and a summer observation. The additional observations comprise 19,074 groundwater connected lakes, 1,000 stream points and 1,000 coastal points. All additional observations were defined with zero depth of the uppermost water table.

95% confidence intervals which can be achieved by training four models, each model targeting a single quantile (e.g., 0.16 and 0.84 for the 68% confidence interval). The four models have to be trained individually for both, summer, and winter condition. We used CPU on a Windows machine (4 2.2 GHz Intel Xeon processors with 56 cores in total, 256 GB RAM) for training and prediction.

RESULTS

High Resolution Groundwater Model

CatBoost regression was used to simulate the depth of the uppermost water table for a typical winter and summer condition at 10 m resolution over entire Denmark. Hyper parameters were first manually calibrated and afterwards automatically fine-tuned using a randomized search approach. Table 3 contains the nine hyper parameters included in the randomized search, a short description, and the tested values. The randomized hyper parameter search was conducted for the summer and the winter model using a 3-fold cross validation approach using 75% of the data. The 3-fold cross validation is the default of CatBoost’s randomized hyper parameter search algorithm. The MAE, based on the 2,500 hyper parameter combination derived from the randomized search algorithm, varied just 4 cm. A common hyper parameter set that was among the top 2% for both, summer, and winter model, was selected for further modeling (Table 3). We found that the performance did not deteriorate for the test data (25%).

Table 4 shows results obtained from a 4-fold cross validation test for the summer and winter models. For this, four CatBoost regression models were trained using 75% of the data for training and 25% of the data was held back for validation. Overall, little variation was found across the cross validation models, which indicates an overall robustness. The MAE, averaged across the four cross validation test, is 47 cm for both, summer, and winter model. This calculation is based on the entire training dataset. Only considering well data yields an average MAE of around 1.15 m. The increase in MAE is due to the fact that the additional observations, all with a depth of zero, are well-captured by the model and generally yield lower residuals. Based on the coefficient of determination over half of the variance in the well data is accounted for by the models and over 70% of the variance of the entire training dataset. The 19,000 lakes are represented with a MAE of below 5 cm, whereas additional observations along the stream network and coastline possess a MAE of around 10 cm.
TABLE 2 | Overview of the explanatory variables used to model the uppermost water table.

| Variable                              | Abbreviation | Description                                                                 | Source                        |
|---------------------------------------|--------------|----------------------------------------------------------------------------|--------------------------------|
| Clay content 0–30 cm                  | ClayA        | Clay content in percentage for four soil layers at 30 m resolution         | Adhikari et al., 2013         |
| Clay content 30–60 cm                 | ClayB        | Clay content in percentage for four soil layers at 30 m resolution         |                                |
| Clay content 60–100 cm                | ClayC        | Clay content in percentage for four soil layers at 30 m resolution         |                                |
| Clay content 100–200 cm               | ClayD        | Clay content in percentage for four soil layers at 30 m resolution         |                                |
| Thickness of top clay                 | ClayThick    | Thickness of the uppermost clay layer at 100 m resolution                   | DK-model                      |
| Landscape typology*                   | LType        | Geomorphological classification in 13 classes as polygon shape file         | Breuning-Madsen and Jensen (1992) |
| Land Use*                             | LUse         | 7 land use classes at 100 m spatial resolution                             | Levin et al. (2012)           |
| Degree of urbanization                | Urban        | Percentage of grid cell that is paved at 10 m spatial resolution           |                                |
| Water Body*                           | WBody        | Binary water layer containing rivers, lakes and coastline at 10 m spatial resolution |                                |
| Elevation model                       | DEM          | Digital elevation model at 10 m spatial resolution                         | The Danish Agency for Data Supply and Efficiency (SDFE) |
| Terrain slope                         | Slope        | Rise and fall of the terrain surface in degree at 10 m spatial resolution  |                                |
| Horizontal distance to WBody          | HDIs         | Horizontal distance to nearest water body at 10 m spatial resolution       |                                |
| Vertical distance to WBody            | VDis         | Vertical distance to nearest water body at 10 m spatial resolution         |                                |
| PBM—winter condition                  | PBMw         | Depth of water table for median winter condition simulated by a PBM at 100 m spatial resolution | DK-model                      |
| PBM—summer condition                  | PBMs         | Depth of water table for median summer condition simulated by a PBM at 100 m spatial resolution |                                |

Categorical variables are indicated with an asterisk.

TABLE 3 | CatBoost hyper parameter used in the randomized search (2,500 combinations).

| Hyperparameter   | Description                                                                 | Tested values     | Optimized value |
|------------------|------------------------------------------------------------------------------|-------------------|-----------------|
| learning_rate    | Reduces the gradient step during training                                    | 0.05, 0.075, 0.1, 0.125, 0.15 | 0.05           |
| Depth            | Number of levels in the decision trees                                       | 8, 9, 10, 11, 12, 13 | 13              |
| l2_leaf_reg      | The coefficient of the L2 regularization term of the loss function           | 0, 2, 4, 6, 8, 10, 12 | 4               |
| Subsample        | Random selection of training data for defining splits                        | 0.5, 0.6, 0.7, 0.8, 0.9, 1 | 1               |
| rsm              | Random selection of covariates for defining splits                           | 0.5, 0.6, 0.7, 0.8, 0.9, 1 | 0.6            |
| random_strength  | Randomness for selecting the optimal split                                   | 0.5, 0.75, 1, 1.25, 1.5 | 0.5            |
| min_data_in_leaf | Minimum data in each leaf                                                    | 1, 5, 9, 13, 17, 21, 25 | 25              |
| bagging_temperature | Random weights to training data                                             | 0, 0.5, 1, 1.5    | 1.5            |

The maximum number of decision trees, i.e., iterations in the gradient boosting, was set to 1,000.

Figure 4 depicts the simulated national maps of the depth of the uppermost water table at 10 m spatial resolution for summer and winter. The maps offer much detail revealing the interplay of geology, topography, and waterbody proximity. The overall regional patterns show resemblance with the simulations results of the physically-based model at 100 m, which were used as covariate in the 10 m ML models (Figure 2). However, disagreements between the 100 m physically-based model and the 10 m ML model are also present. The former simulates a shallower water table in the Eastern part of Denmark where moraine clay is the dominant lithology. This is especially the case for the winter condition. Further, the 100 m physically-based model simulates a homogeneously deep water table in the sandy areas whereas the ML model results in more heterogeneity. The disagreements may be explained by the differences in spatial resolution, but also with the vertical discretization of the computational layers in the physically-based model. Sandy layers quickly run dry which results in a deep water table and clayey layers hold water which yields a very shallow water table. The ML based median seasonality is calculated as the difference between the summer and the winter maps. The amplitude is generally lowest close to streams and lakes and highest in the center of
TABLE 4 | Results for a 4-fold cross validation (cv) applied on the summer and winter model.

|                   | Summer model | Winter model |
|-------------------|--------------|--------------|
|                   | cv1  | cv2  | cv3  | cv4  | Mean | cv1  | cv2  | cv3  | cv4  | Mean |
| All data          | MAE  | 0.48 | 0.48 | 0.48 | 0.46 | 0.47 | 0.47 | 0.47 | 0.45 | 0.47 |
|                   | RMSE | 1.24 | 1.21 | 1.23 | 1.19 | 1.22 | 1.23 | 1.21 | 1.18 | 1.20 |
|                   | $R^2$ | 0.75 | 0.76 | 0.75 | 0.77 | 0.76 | 0.72 | 0.72 | 0.71 | 0.73 |
| Only well observations | MAE  | 1.17 | 1.18 | 1.19 | 1.15 | 1.17 | 1.16 | 1.17 | 1.18 | 1.16 |
|                   | RMSE | 1.96 | 1.92 | 1.95 | 1.91 | 1.93 | 1.94 | 1.92 | 1.93 | 1.92 |
|                   | $R^2$ | 0.51 | 0.51 | 0.51 | 0.54 | 0.52 | 0.51 | 0.52 | 0.51 | 0.51 |
| Only lakes        | MAE  | 0.03 | 0.04 | 0.04 | 0.03 | 0.04 | 0.03 | 0.03 | 0.03 | 0.03 |
|                   | RMSE | 0.30 | 0.27 | 0.33 | 0.27 | 0.29 | 0.26 | 0.25 | 0.28 | 0.23 |
|                   | $R^2$ | 0.16 | 0.16 | 0.16 | 0.16 | 0.16 | 0.16 | 0.16 | 0.16 | 0.16 |
| Only river and coastline | MAE  | 0.13 | 0.11 | 0.13 | 0.12 | 0.12 | 0.09 | 0.08 | 0.11 | 0.10 |
|                   | RMSE | 0.31 | 0.31 | 0.35 | 0.32 | 0.32 | 0.27 | 0.22 | 0.28 | 0.27 |
|                   | $R^2$ | 0.51 | 0.51 | 0.51 | 0.51 | 0.51 | 0.51 | 0.51 | 0.51 | 0.51 |

Performance is quantified by means of a mean absolute error (MAE), root mean squared error (RMSE) and coefficient of determination ($R^2$). Results are given for different subset of the data (all data, only wells, only lakes and only river and coastline).

Denmark where topography is high and a permeable subsurface is present. Similar to the physically-based simulated amplitude, the ML derived amplitude has intermediate values around 0.5 m (yellow category) in the Western part of Denmark for the sandy outwash plains. Based on both approaches, the amplitude for the moraine clay settings is around 1 m (green category) which is mainly driven by the very shallow water table during winter.

Figure 5 presents the same data as shown in Figure 4, just for a zoom section of ~15 km$^2$. Here, the imprint of the stream network as well as the many lakes, become apparent as areas with a very shallow depth of the uppermost water table (<0.5 m). The difference between the drier summer and the wetter winter is clearly notable and the difference between the two models is plotted as the seasonality. The largest amplitude is present in areas with high topography and the lowest amplitude is found along the river network and lakes.

Covariate Importance

The quantified importance of each input feature for the summer and winter model is presented in Figure 6 for the entire training dataset and for a subset containing only the well data. Based on a trained model, CatBoost calculates the “prediction value change” to quantify how much on average the prediction changes if the covariate value changes. The average changes that represent the importance of a given covariate are normalized to add up to 100. The vertical distance to the closest water body (VDis) clearly stands out as the most importance covariate in both models. At locations with VDis close to zero, the water table is typically also close to the surface. However, large VDis values, which indicate small scale topographical variations often result in a deeper water level, as the shallow groundwater does not follow the topography in such settings. The most important categorical covariate is the landscape type classification (LType) which contains 13 landscape classes, such as moraine, marine plains, outwash plains, and others. There is little difference between winter and summer model, which indicates robustness between the two models. Differences between covariate importance with respect to the entire dataset and only well data conveys that the horizontal distance to the closest water body (HDis) is mostly relevant to the additional observations, namely lakes, rivers, and coast, which is expectable as these are characterized with a distance of zero. The physically-based model (PBM) stands out as second most important covariate when only considering the well-training dataset. This underlines that the physically-based model can provide the ML model with a meaningful information, despite the differences in spatial resolution, since both model the same variable.

Uncertainty Analysis

Quantile regression has been applied in order to estimate the uncertainties associated with the summer and winter models. For both models, four quantile models have been trained using the same hyper parameters as applied previously. The quantile models were set with $q = 0.16$ and $0.84$ for the 68% confidence interval and $q = 0.025$ and $0.975$ for the 95% confidence interval. Results are shown in Figure 7 and were calculated on the basis of the same 4-fold cross validation test as presented in Table 4. Figure 7 only contains wells and data are sorted with respect to the simulated groundwater depth and increase alongside an increasing x-axis. The observations, 13,047 in total, are plotted as a density plot and an overall good agreement between model and observations can be attested to both, summer, and winter. The blue envelope plots indicate the two confidence intervals and following the quantile regression definition, 5% of the observations are expected to be outside the dark blue envelope (95% confidence) and 32% are expected to be outside the dark blue envelope (68% confidence). For both, the summer, and winter model, the uncertainty increases with depth, which gets supported by the large spread of observations for larger depths.
Groundwater depths above 6 m have an uncertainty of around 4 and 8 m following the 68 and 95% confidence intervals.

Based on the 95% confidence interval, which reflects the spread of $\sim \pm 2$ standard deviations around the mean, a map of the standard deviation at 10 m spatial resolution has been calculated. Overall, the relationship between the standard deviation and the water table depth is nearly linear and the coefficient of determination, expressed as the average standard deviation over the average water table depth is 0.58. This indicates that the uncertainty is roughly half of its water table depth value. Figure 8 exemplifies the relationship between the simulated depth of water table and the associated uncertainty (standard deviation) for two soil types. In moraine clay soils, the depth of the uppermost water table is typically in the top few meters, whereas moraine sand soils are characterized with deeper water tables. The coefficient of variation for the two soil types is 0.74 for moraine clay and 0.44 for moraine sand which underlines that uncertainty is larger for moraine clay soils as opposed to moraine sand soils. This can be expected given the distinct geological heterogeneity in the moraine clay soils.

**DISCUSSION**

**Training Dataset**

The winter trainings dataset comprises all DJF water level observations and JJA represents summer conditions. The median was calculated in the case of multiple observations per well per season. This approach introduces uncertainties since it rules out inter-annual variability and further, variability within the summer and winter seasons is also ignored. This compromise was accepted in order to obtain a large trainings dataset of groundwater observations. Large-scale groundwater datasets are typically very heterogenous, which is especially the case for the temporal dimension. The spatial density of wells may be high, but the temporal resolution of water level observations poses challenges to ML applications. Heterogeneity originates...
from varying frequencies and periods of observations. As an example, around 66% of the 104,000 wells in Denmark have only a single observation in the period of 1990–2019. In order to capitalize on under sampled shallow wells in a big data context, this study brings forward a knowledge-guided ML framework that employs the simulated seasonality from a physically-based groundwater flow model. With this augmentation strategy, the training dataset could be expanded significantly from 1,378 shallow wells with both, summer and winter observations, to 13,047 with either gap-filled summer or winter observation. Previously, Koch et al. (2019b) employed a temporal projection of any given well observation to a representative extreme winter condition using a simple sinusoidal model that was fitted to numerous hydrogeological units. While this approach was purely observational based it was limited to a few pre-defined hydrogeological units whereas the knowledge-guided framework presented herein made use of a spatially distributed groundwater model that provides groundwater amplitudes, i.e., seasonality, at higher spatial detail. Further, the training dataset was extended by groundwater proxy observations featuring lakes, rivers, and coastline where the water table depth is expected to be zero. These information were partly derived from expert knowledge, which is another important aspect of the concept of knowledge-guided ML, which has special relevance for the data-scarce groundwater domain. The simulation results of the physically-based groundwater model were employed to identify wells that are representative for the shallow groundwater system, i.e., representing the uppermost water table, by means of soil type dependent threshold depths. This process step was essential to train the ML against a homogenized well dataset and omitting this would result in a dataset with well observations containing water levels of different aquifer systems.

This study utilized a comprehensive set of 15 covariates to predict water table variability. This selection has been guided by a previous Danish study by Koch et al. (2019b). As an additional covariate, satellite-based radar data, which have shown to be sensitive to shallow water table dynamics within the first couple of meters below surface (Bechtold et al., 2018), could potentially be employed as a meaningful input to the future developments of the framework presented in this study. Other high resolution satellite data on land surface temperature or shortwave infrared (obtained from e.g., Landsat) could provide relevant information to differentiate moisture conditions in areas where land use is constant and the water table is close to the surface.

**Temporal Resolution**

This study implements a simplified temporal dimension of the groundwater table, by simulating two seasons, namely a typical winter and summer condition. These temporal snapshots are of course a prude simplification of a variable that is known to possess a distinct temporal variability. However, previous studies that apply ML to model the spatial variability of the water table have focused on a single time step; i.e., Koch et al. (2019b) model an extreme wintertime condition and Fienen et al. (2013) and Bechtold et al. (2014) focus on mean annual conditions. ML applications that focus on time series modeling of the water table are typically limited to a few wells with complete time series of daily or monthly observations spanning over several years (Sun et al., 2016; Guzman et al., 2017; Wunsch et al., 2021). Such modeling techniques, which are often based on deep learning due to its strengths in sequence modeling, have not yet been rolled out to model the full spatio-temporal variability of groundwater dynamics due to geological complexity and data scarcity. Full spatio-temporal modeling schemes would require thorough testing on how a dynamic ML model can be transferred from one location to another. This is especially challenging for the groundwater system as the temporal water table variability depends on the geology and the three-dimensional connectivity of the subsurface, which is often poorly described by available data sources at larger scales. This limitation favors the knowledge-guided ML approach where a physically-based groundwater model is utilized to integrate all subsurface data in a meaningful way to inform a ML model on the effect of subsurface connectivity and geology on water table dynamics.

**Machine Learning Model**

Overall, the accuracy of the trained ML model, which was quantified by means of a 4-fold cross validation test, was very satisfying and in the range of what is generally considered very acceptable in groundwater flow modeling (Henriksen et al., 2003). This underpins the applicability of gradient boosting decision tree models to adequately represent complex, non-linear variables, such as water table depth. The applied resulting spatial resolution of 10 m provides an insightful screening tool for water management purposes, such as the risk assessment of groundwater floods on agricultural fields or urban areas, which are both very relevant for Danish applications. Such flooding is typically very local and driven by small-scale variations of topography and geology, which emphasizes the need for high-resolution predictions to reliably tackle the related challenges. At the national scale, a spatial resolution of 10 m resolution would not be feasible with conventional numerical modeling.
tools, which promotes the versatile applicability of ML for the groundwater domain. We believe that the input to the trained ML model supports the applicability of the models at 10 m resolution. The covariate importance attested high importance to the topography related covariates, which have a native resolution of 40 cm and thereby supports the 10 m resolution of the ML model. This may only holds for the uppermost water table. The given covariates will likely not support a 10 m ML model of deeper aquifers, that are less topography controlled.

Results from a physically-based groundwater model were incorporated in the ML model as explanatory variable and it was shown that the importance of the 100 m groundwater model was the second highest with respect to the well observations. This is satisfying as it issues consistency between the two modeling approaches and underpins that the physically-based model can guide the ML model. It has to be noted that the wells used for training the ML model were also used to calibrate the groundwater model. However, we believe that reusing of data does not have any implications and in the end, a resemblance between the groundwater model and the ML model is also desirable. The most important explanatory variable is the vertical distance to the nearest waterbody, which indicates that the shallow groundwater system is decoupled from small scale topographical variability. The applied knowledge-guided ML framework is novel for the groundwater domain and may be applicable to other regions where results from a physically-based groundwater model are available. Adding simulation results from hydrological models into ML models has also been successfully applied for the surface water system, i.e., predicting streamflow (Konapala et al., 2020). There certainly exists various levels of knowledge-guide ML strategies and augmenting data with simulation data maybe considered the initial step. More advanced setups may incorporate first principles or partial differential equations that eventually alleviate data scarcity constraints and increase interpretability of ML models.

We envision that the developed water table map can support the planning of climate resilient infrastructure design, with special focus on flooding or the planning of rewetting of lowlands to reduce greenhouse gas emissions from drained peatland soils. In contrast to conventional PBMs, the ML proposed herein cannot be used to run modeling scenarios (e.g., climate change or water management), but our results can be used as a national scale screening tool to identify areas where a PBM can subsequently be applied to test relevant scenarios at high spatial resolution. The ML based results reflect the current climate conditions and if the impact of climate change of the uppermost water table requires investigation, a PBM should be applied instead.

**Uncertainty Analysis**

The quantile regression technique was used to estimate uncertainty bounds. Uncertainty was quantified corresponding to the 68 and 95% confidence intervals of the simulated water table depth. We found that uncertainty generally increases with depth and that the average coefficient of variation is 0.58, indicating that the standard deviation is more than half of the water table depth, but depends on the geological setting with higher uncertainty for the complex moraine clay soils. One known drawback of quantile regression is that it requires individual training of each selected quantile, which can result in an invalid distribution, meaning that the estimated quantiles are not monotonically increasing (Bondell et al., 2010). This problem is often referred to as “crossing quantiles” and different strategies to overcome this limitation were proposed by López López et al. (2014). Typically, crossing qualities occur more frequently close to the median. For the estimated water level quantiles in this study we found that by selecting the 95% confidence intervals ($q = 0.025$ and $0.975$) crossing quantiles were not occurred, whereas this was partly the case for the 68% ($q = 0.16$ and 0.84) confidence intervals.

Following the discussion provided by Vaysse and Lagacherie (2017), quantile regression addresses where a prediction point is located in the covariate space and how well it is constrained by the available observations. In this way, quantile regression can properly discriminate water level conditions of contrasted physical complexities, of which some are better represented by the training dataset than others. For the shallow groundwater system, uncertainties could originate from the actual water level observations in the wells, the augmentation strategy using a physically-based groundwater model to gap-fill missing summer or winter observations, or the fact that the training dataset is processed for a 30-year period and inter-annual variability has been ignored. In contrast, a physically-based groundwater model allows more transparency, as erroneous
Koch et al. High Resolution Water Table Model

FIGURE 8 | Density scatter plot showing the simulated depth of the groundwater for moraine clay and sand soil against the associated standard deviation (std.) for the winter model. Results are shown for two of the nine soil types (Figure 1).

water-level observations will be marked as outliers in the model evaluation. However, a flexible data-driven model will incorporate such outliers as it seeks an unbiased representation of all observations, which emphasizes the need for a careful selection and pre-processing of the training data. Ultimately, uncertainty can be reduced by expanding the training dataset with high quality additional water level observations. Additional explanatory variables could be an alternative strategy to reduce uncertainty.

CONCLUSIONS

The study features a knowledge-guided ML model of the uppermost water level at 10 m spatial resolution at national scale for Denmark. We have applied the gradient boosting decision tree implementation of CatBoost to model a typical summer and winter condition. The associated uncertainties were estimated using quantile regression techniques. Predicting water levels at 430 million grids is unfeasible with conventional dynamic physically-based groundwater flow models, which highlights the benefits of using alternative ML modeling approaches instead, to reach unprecedented spatial detail. We draw the following main conclusions from our work:

- The applied high resolution ML model could predict water table variability with high accuracy. The MAE of well observations was around 115 and 50 cm taking also the groundwater proxy observations (lakes, rivers, and coastline) into consideration.
- A physically-based groundwater flow model was successfully incorporated into the model building to (1) select wells that are representative for the shallow groundwater system, (2) augment training data by accounting for simulated water level seasonality, and (3) extend the list of explanatory variables. This forms a novel application of knowledge guided ML for the shallow groundwater domain.
- The water table depth was simulated for two temporal snapshots (typical summer and winter). Future ML research in the groundwater domain must focus on modeling the full spatio-temporal variability of water table depth.

DATA AVAILABILITY STATEMENT

Publicly available datasets were analyzed in this study. The water table observations are available via the Jupiter database: http://data.geus.dk/geusmap. The raw data and scripts supporting the results and conclusions of this article will be made available by the corresponding author, without undue reservation. The final summer and winter water table depth maps are visualized and freely available via https://hipdata.dk/.

AUTHOR CONTRIBUTIONS

JG: data curation. JG, RS, LT, SS, and HH: conceptualization, result interpretation, and quality control. JK: code development, study design, writing original draft, and visualization. All authors have read, edited, and agreed to the published version of the manuscript.

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