Scots pine’s capacity to adapt to climate change in hemi-boreal forests in relation to dominating tree increment and site condition

Marius Mikalajūnas (1), Hans Pretzsch (2), Gintautas Mozgeris (1), Edgaras Linkevičius (1), Ingrida Augustaitienė (1), Algirdas Augustaitis (1)

Forest site (FS) and meteorological conditions are recognized as the main factors affecting tree growth and whole-stand sustainability. This study aims to detect the combined effects of FS and meteorological conditions on tree ring formation of Scots pine (Pinus sylvestris L.), the most common tree species in Lithuania and hemi-boreal forests of northeastern Europe. We used data on stand structure and productivity from the Lithuanian National Forest Inventory (NFI) and stem radial increment series of dominating trees during the period 1993-2012 collected since 2013. Pine stem basal area increment (BAI) was chosen as the response variable, while temperature in March (°C) and precipitation in June (mm) were used as predictor variables, as they best express the effect of climate change on Lithuanian forests. We simulated the effects on dominating pine annual increment of deciduous tree species, mainly Betula sp. and the level of soil moisture and fertility, accounting in addition for the random effects of NFI network tract, plot direction, and tree number. A nonlinear mixed-effects model explained up to 68% of the variation in the BAI of pine trees. The annual pine trees BAI increased with the increase in the proportion of deciduous trees in pine stands. Increases in temperature and precipitation in considered months reinforced this positive effect on pine BAI, especially in mature pine stands in temporarily waterlogged meso-eutrophic FSs on mineral soils. A negative effect of deciduous trees on pine stem increment was observed only in nutrient-rich eutrophic and drained peatland FSs. Forestry treatments directed towards the increase in deciduous tree proportion in the most common normal or temporarily waterlogged meso-eutrophic and oligotrophic pine stands might increase the biodiversity and productivity of pine stands, and their sustainability in future climate change scenarios.

Keywords: Scots Pine, Basal Area Increment, Site Conditions, Meteorology, Mixed-effects, Hemi-Boreal Forests

Introduction
There is great concern worldwide about the effects of climate change on forest growth, because forests account for a large portion of the sequestered carbon by terrestrial ecosystems. Climate change is expected to increasingly impact forest ecosystems (Bouwman et al. 2021), as it could affect forest area, health and biodiversity. On one hand, climate change may increase growth rates in temperate and boreal forests (Mensah et al. 2021), while on the other hand, temperature-induced drought stress can endanger the survival of trees and forest communities (Pretzsch et al. 2013, Toochi 2017, Pretzsch 2020), especially in southernmost areas (Rubio-Cuadrado et al. 2020). In northeastern Europe, such changes are expected to have a positive influence on boreal forest growth with an increase in productivity. However, the impact of climate warming in the hemiboreal zone, which is the transition zone between boreal and temperate zones, depends on site conditions (Matisons et al. 2021). Such divergent forest productivity trends challenge our capability to effect projections of tree growth dynamics under the pressures of climate change (Ols et al. 2019).

The impact of climate change on productivity varies according to geographic area, species, stand composition, tree age, soil fertility and water holding capacity, as well as the interactions of these parameters (McMillan et al. 2008, Reich & Oleksyn 2008, Ols et al. 2019). Therefore, forest site-specific conditions like soil moisture and fertility can be key factors limiting tree growth (Forrester & Albrecht 2014). However, our understanding of the effects of climate on forest health, productivity, and the carbon cycle is limited for many regions and ecosystems, especially under different or opposite site conditions (Spiecker 1999, Nellmann & Thomsen 2001, Chapagain & Sharma 2021). Usually, such observations are based on small sample sizes, which makes it difficult to extrapolate them to the regional or national forest scale (Sharma et al. 2019). The influence of species mixing on the reaction of pine to environmental change needs more detailed investigation, especially in different site conditions (Aldea et al. 2021, Chapagain & Sharma 2021).

Tree responses to climate change have been analyzed in relation to soil conditions (Reich & Oleksyn 2008). Tree rings are widely used to study the impact of climate change on forest carbon cycling, and for validating process-based models of forest productivity where temporal variation are determined by large trees (Xu et al. 2019).
Understanding the effects of climate change on forest productivity and its time series is critical, and tree rings offer an annually resolved proxy to explore tree growth over times when no forest inventories or meteorological instrumental records were available (Cook & Kairiukstis 1990). Tree-ring series are widely used to reconstruct stand productivity history. Therefore, the impact of climate change on forest tree growth, as well as past growth trends (Macias et al. 2006, Piovesan et al. 2008) have been extensively analyzed through series of tree rings (Girardin et al. 2008, Garant et al. 2009), especially of dominating trees in forest stands (Xu et al. 2019).

Scots pine (Pinus sylvestris L.) was chosen in this study as the main species to detect the effect of increased temperature and precipitation on tree annual increment. It is one of the most common tree species not only in Lithuania but also in the entire boreal and boreal European region (Pretzsch et al. 2015). Our earlier findings in northeastern Lithuania revealed that since 1980 Scots pine trees showed significant increases in annual basal area increment (BAI). This was related to significant long-term increases in air temperature in September, early spring and July. Heat and drought in June limited pine tree growth, but only on organic soils under excess-moisture regime, which increased the sensitivity of pine to drought (Augustaitis 2007a, 2015, 2018).

In northeastern Europe, air temperature and precipitation increased over the 20th century and particularly since the late 1980s (IPCC 2014, Augustaitis et al. 2018, Ols et al. 2019). Recently Lithuania became a top hotspot in Europe. There greatest shifts in thermal and precipitation regimes have been detected, with mean temperature increasing by +0.325 °C and precipitation by +20 mm per decade (Schlechtriem 2019). The adaptation capacity of pine forests to such climate changes is one of the major challenges for forest management in this region. It is necessary to understand the effects of climate change on pine growth to be able to predict how these effects might affect pine stand sustainability in boreal forests in the future. Recently, the proportion of deciduous tree, especially Betula spp. in pine stands, has also been shown to have a major effect on pine tree annual increment in relation to tree age, density and site condition (Aldea et al. 2021). In this study we attempted to determine and assess: (i) the effects of key meteorological parameters on pine tree annual BAI in forest sites with different levels of soil moisture and soil fertility; (ii) the effect of the proportion of deciduous trees in pine stands on pine annual BAI under different soil conditions; (iii) the effect of tree age on pine tree annual BAI under different forest site conditions in relation to selected meteorological parameters; (iv) future changes in pine tree annual BAI in boreal pine forests under forecasted climatic conditions.

The results could suggest new forest management treatments ensuring sustainable forest development under the pressures of global change in boreal forests.

**Materials and methods**

**National Forest Inventory data**

Data on stand, tree and soil characteristics were collected on the NFI plots during the fourth National Forest Inventory (NFI) 5-year cycle from 2013 to 2017 (Fig. 1). Scots pine cores were collected at breast height from one selected dominating tree of each tree species growing in the same stand, but outside of the permanently monitored plot. Annual increment of the dominating tree of each tree species in a stand was selected as the main response variable to detect the integrated effects of environmental predictor variables. The sample size for dendrochronological analysis during the entire 5-year cycle consisted of 2414 Scots pine trees over 40 years old from the same number of sample plots. More than 2500 cores were taken from other tree species. We eliminated data on pine BAI from stands younger than 40 years to reduce variability in tree-ring width data, as the variability in younger stands is high due to tree competition and tree density (Kairiukstis et al. 1987, Schweingruber 1996, Jukny et al. 2002, 2003, De Vries et al. 2014).

Dendrometric parameters of trees were obtained from four circle plots (500 m² in size and 12.76 m in radius) in each NFI tract, which were located on the tops of isosceles triangles at 4 km distance from each other both horizontally and vertically (Fig. 1). The identification of a tract and the position of a plot in the tract (direction), as well as the tree identification number (id) were used to account for random effects on pine BAI formation in mixed-effects models.

Forest type and soil conditions were also assessed in the plots using the Lithuanian NFI methodology (Kuliešis et al. 2016).

**BAI calculation of sample dominating trees**

We measured annual tree-ring width in stem cores, from bark to pith, with an accuracy of 0.01 mm, using an electronic transducer and a binocular scope fixed over the moving stage of the Lintab v.6 tree-ring measuring equipment, and TSAP-Win™ software (RinnTech e.K., Heidelberg, Germany). Individual tree-ring width series were synchronized by a visual comparison of graphs for each ring width series with averaged pine tree ring width series of the correct age group (Eckstein et al. 1989). We calculated the Pearson’s correlation coefficients and means among and between each tree ring width series for the right age group (Baillie & Pilcher 1973). This procedure ensured significant differences in annual increments among years, but synchronous and non-significant differences in the rhythm of growth in relation to tree diameter at different ages (Pretzsch 2020).

The data on tree ring widths over the last 5-year period (2013-2017) were not included in the analysis to avoid not fully formed increments of the last year to have equal period for each tree.

To detect recent regular pine tree

---

**Fig. 1** - Location of Scots pine sample plots and structure of NFI permanent plots at 4 km distance from each other across Lithuania. The 250 × 250 m tract with four 500 m² circle plots at its different directions represents a plot.
growth, the period between 1993 and 2012 was chosen for analysis (Fig. 2). This 20-year period when mean temperature and precipitation increased by 0.049 °C and 1.92 mm per year, respectively, (Augustaitis et al. 2018) reflects climate change quite well. This increase is statistically significant and is in full agreement with the SRES A1B Project (IPCC 2014).

Basal area increment (BAI) series were derived from tree ring width series. BAI was used instead of ring width, due to its higher relationships with stem volume increment and lower interrelationship with tree diameter and tree age processing, as well as recent tree age. This parameter captures growth trends better than tree-ring width and avoids the need for data detrending (Biondi & Qeadan 2008). Such detrending could also remove low frequency variability that could hide the effects of meteorological parameters with the same variability.

Meteorological data

Meteorological data were collected from 16 meteorological stations located across Lithuania. Mean monthly precipitation and temperature were used in the BAI analysis. Mean temperature in March and precipitation during June were chosen as they have the most statistically significant differences between the effect of predicting variables on pine BAI in the main forest site and the rest FSs, while p>0.05 indicated no statistical differences of this effect, i.e., effect of predicting variables was the same like in the main selected forest site. BAI model in relation to soil fertility level of FS is described as follows (Nutrient model – eqn. 2):

$$\log(\text{BAI}) = \log(\text{Water}) - \log(\text{Mix}) \cdot \log(\text{BA})$$

where BAI is the basal area increment in year t. Fixed effects are composed of four “Nutrients” groups based on soil fertility level (Nutrients: N1, N2, N3, N4); T in March; P in June; Mix is the proportion of pine trees in the mixed pine stand; BA is the tree basal area in year t-1, calculated from ring width.

An interaction comparison was performed on normal moisture sites (W1), which were the most common with N = 1682 trees. Where p<0.05 indicated significant differences between the effect of predicting variables on pine BAI in the main forest site and the rest FSs, while p>0.05 indicated no statistical differences of this effect, i.e., effect of predicting variables was the same like in the main selected forest site.

Scots pine BAI modeling

The linear mixed-effects (LMe) model from the R package “NLME” (Pinheiro et al. 2018) was used to analyze differences in annual BAI of different soil conditions. Tract ID, plot position in the tract, and tree ID were included as random effects in all the models. An asterisk (*) in the model indicates interactions between the main group and the remaining three “Water” or “Nutrients” groups. The predict function in the R package “stats” v. 3.6.0 was used to display the model values with average group parameters.

BAI data logarithms (log) were used to remove heteroscedasticity from the model residuals and after the simulation were back-transformed to tree BAI using exponent function. Biological plausibility of the results was used to evaluate the model (Burnham & Anderson 2002). All the statistical analyses were carried out using R software (R Development Core Team 2018).

The BAI model in relation to the moisture regime of the FS referred to as “Water” is described as follows (Water model – eqn. 1):

$$\log(\text{BAI}) = \log(\text{Water}) \cdot (T_{\text{in}} + P_{\text{in}} + \log(\text{Mix})) \cdot \log(\text{BA})$$

where BAI is the basal area increment in year t. Fixed effects are composed of four soil moisture regime groups (Water: W1, W2, W3, W4); T in March; P in June; Mix is the proportion of pine trees in the mixed pine stand, where 1 indicates a pure pine stand; and BA is the tree basal area in year t-1, calculated from ring width.

Interactions between BAI and predict variables were compared to the mesotrophic site (N2) as this FS is not only quite represented in our sample, but is also the FS where the most productive pure pine stands could be developed in hemi-boreal coniferous forest.

In the present study, the effect of deciduous trees in pine stands is expressed as the proportion of pine trees in the pine stand (“Mix”), defined as (eqn. 3):

$$\text{Mix} = 1 - \frac{\text{DBA}}{\text{WSBA}}$$

where DBA is the basal area of deciduous trees in a pine stand; WSBA is the whole...
stand basal area. To evaluate the effect of tree age on the relationships between pine BAI formation and meteorological variables in different mixed pine stands, data obtained in normal humid (W1), temporary waterlogged (W2), mesotrophic (N2), and oligotrophic (N3) soil FSs were used. Very oligotrophic (N4) FS stands were excluded from this analysis due to the inadequacy of oligotrophic soils for the growth of deciduous trees. Three age groups were selected, each age group containing more than 200 samples of pine BAI series from the same pine stands. BAI model in relation to the selected tree age, referred to as “Age”, is described as follows (Age model – eqn. 4):

$$\log(\text{BAI}) = (\text{Age} - T_{\text{min}} P_{\text{W1}} + \log(\text{Mix})) + \log(\text{BA})$$

where the fixed effects are composed of Age, classified into three age groups (60-, 80-, and 110-year-old); $T$ and $P$ parameters, which are as defined above for the “Nutrients” and “Water” models. An interaction comparison was performed on the youngest 60-year age group.

Determination coefficients ($R^2$) were used to evaluate the models. Marginal $R^2$ describes the proportion of variance explained by the fixed factors alone; conditional $R^2$ describes the proportion of variance explained by the fixed and random factors (Nakagawa & Schielzeth 2013). The effects of fixed factors in the BAI model were evaluated using the level of significance ($p$). Model estimate values were back-transformed using the "exp" function in R.

Results

Overall data evaluation

The sampled trees for BAI estimation were distributed in different FS groups based on “Nutrients” and “Water” levels. The regular FS condition of pine stands in Lithuania is normal humid (W1) and oligotrophic soil (N3). These groups were the most represented in the study. Representation of the other FSs was lower (Fig. 3). Pine tree stands from drained peatland (W3) and eutrophic soil (N1) FSs were less represented in the study, because these FSs are less represented overall for pine growth in Lithuania. The most productive FSs were on temporarily waterlogged soils (W2), and mesoeutrophic soils FS (N2). BAI data reflect the productivity of the separate FS groups. The least productive pine stands grow on undrained peatland (W4) and very oligotrophic (N4) soils. Pine growth at these sites is extremely limited by an over-moisture regime and lack of nutrients compared to the other FSs. Only a few sampled pine trees represented eutrophic soil (N1) FSs, as these sites are mainly occupied by deciduous tree species and the pines growing there tend to form big crowns with large branches that ruin the overall wood quality and do not form pure stands. This FS is not typical of pine growth in a hemi-boreal forest.

The results showed that temperature in March had a significant effect on pine BAI variation at W1 forest sites (Tab. 2 – Water model, MT row). March temperature also had a positive effect on the other FSs. An increase in temperature by 1 °C resulted in an increase of pine BAI in the different “Water” FS groups of 0.19 cm² for W1, 0.26 cm² for W2, 0.26 cm² for W3, and 0.07 cm² for W4 (Fig. S1 in Supplementary material). The only significant difference was found between the effect of temperature on BAI in temporarily waterlogged (W2) vs. normal moisture (W1) FSs (Tab. 2 – W2:MT, $p < 0.001$).

Pine BAI in drained peatlands (W3) and natural peatlands (W4) differed also from the normal moisture sites (W1) when the effect of temperature on BAI formation (Tab. 2) was accounted for, but only close to the level of significance ($p < 0.1$ – Fig. S1 in Supplementary material).

The effect of June precipitation on pine BAI formation at W1 FS was also highly significant (Tab. 2 – Water model, JP row), but less significant than the effect of March temperature in other FSs. In W1, W2 and W4 FSs, a 10 mm increase in precipitation increased BAI by 0.16 cm², 0.32 cm² and 0.05 cm², respectively. Precipitation in drained peatland soils (W3) had a negative effect on the BAI of -0.26 cm² per 10 mm, significantly different from the effect of

---

**Tab. 1** - Mean stand characteristics of FSs in soil water and nutrients groups, and their core samples. (BAI): annual basal area increment for a 20 year period from 1993 to 2012 (cm²); (Age): stand age (year); (Mix): mean proportion of pine in the mixed pine-deciduous stand; (Dbh): mean diameter of trees at breast height (1.3 m) at a sample plot (cm); (H): mean height of trees at a sample plot (m); (Age): age of the stand (years); (Cores): number of sampled trees selected for BAI estimation (units).

| Data set | Mean | Unit | Water | Nutrients |
|----------|------|------|-------|-----------|
| NFI      |      |      |       |           |
| Dbh      | cm   | 28.9 | 29.2  | 24.4      | 14.8 | 14.8 | 29.4 | 30.2 | 28.1 | 21.3 |
| H        | m    | 25.9 | 26.2  | 23.1      | 14.1 | 26.1 | 27.1 | 25.4 | 18.5 |
| Mix      |      | 0.95 | 0.88  | 0.84      | 0.91 | 0.80 | 0.88 | 0.94 | 0.97 |
| Age yrs  |      | 85   | 85    | 90        | 88  | 83   | 81   | 87   | 85   |
| Cores    | n    | 1682 | 286   | 139      | 157 | 37   | 471  | 1464 | 292  |

**Tab. 2** - Mean characteristics from dominant in stand trees selected for BAI estimation. (*): Mean characteristics of pine stands obtained from NFI data base; (**): Mean characteristics from dominant in stand trees selected for BAI estimation.

| Data set | Mean | Unit | Water | Nutrients |
|----------|------|------|-------|-----------|
| NFI      |      |      |       |           |
| BAI      | cm²  | 10.36| 10.98 | 9.12     | 4.91 | 12.38| 12.21| 9.54  | 7.71  |
| Dbh      | cm   | 32.3 | 34.0  | 29.1     | 19.3 | 37.0 | 35.3 | 31.4  | 24.9  |
| Age yrs  |      | 77   | 78    | 81       | 87  | 74   | 74   | 80    | 77    |
| Cores    | n    | 1682 | 286   | 139      | 157 | 37   | 471  | 1464  | 292   |

**Fig. 3** - Annual basal area increment (BAI) of the sampled dominating pine trees in pine stands of different moisture and fertility levels. Shaded area indicates the 20-year period of investigation.
Tab. 2 - Models’ parameters and significance values of predicting variables. NFI tract, direction, and tree id were used as random effects. (MT): mean temperature in March; (JP): precipitation amount in June; (SD): standard deviation of the random effects. For more details on parameters, see text.

| Model       | Parameter | Estimate | Std. Error | p-value |
|-------------|-----------|----------|------------|---------|
| Water model | Intercept | -0.124   | 1.053      | <0.001  |
|             | W2        | -0.919   | 0.089      |         |
|             | W3        | 1.134    | 0.101      |         |
|             | W4        | -0.71    | 0.001      |         |
|             | MT        | 1.022    | 1.101      | <0.001  |
|             | JP        | 1        | 0.001      |         |
|             | log(Mix)  | -0.701   | 0.005      |         |
|             | log(BA)   | 3.687    | 0.001      |         |
|             | W2:MT     | 1.007    | 0.014      |         |
|             | W3:MT     | 1.004    | 0.071      |         |
|             | W4:MT     | -0.996   | 0.069      |         |
|             | W2:JP     | 1        | 0.186      |         |
|             | W3:JP     | -0.999   | 0.003      |         |
|             | W4:JP     | -0.999   | 0.769      |         |
|             | W2:log(Mix)| -0.819 | 0.381      |         |
|             | W3:log(Mix)| 2.851 | 0.001      |         |
|             | W4:log(Mix)| -0.62   | 0.273      |         |
| Nutrients model | Intercept | -0.106   | 0.112      | <0.001  |
|             | N1        | 1.29     | 0.106      |         |
|             | N3        | -0.999   | 0.99       |         |
|             | N4        | -0.986   | 0.803      |         |
|             | MT        | 1.019    | 1.001      | <0.001  |
|             | JP        | 1.001    | 0.001      |         |
|             | log(Mix)  | -0.694   | 0.018      |         |
|             | log(BA)   | 3.84     | 0.001      |         |
|             | N1:MT     | 1.003    | 0.568      |         |
|             | N3:MT     | 1.005    | 0.001      |         |
|             | N4:MT     | 1.003    | 0.114      |         |
|             | N1:JP     | -0.999   | 0.592      |         |
|             | N3:JP     | -0.999   | <0.001     |         |
|             | N4:JP     | -0.999   | 0.002      |         |
|             | N1:log(Mix)| 2.509 | 0.072      |         |
|             | N3:log(Mix)| 1.245 | 0.277      |         |
|             | N4:log(Mix)| 1.343 | 0.507      |         |
| Age model   | Intercept | -0.088   | 0.001      | <0.001  |
|             | 80        | 0.964    | 0.582      |         |
|             | 110       | -0.841   | 0.011      |         |
|             | MT        | 1.018    | 0.001      |         |
|             | JP        | 1.001    | 0.001      |         |
|             | log(Mix)  | -0.565   | 0.014      |         |
|             | log(BA)   | 4.109    | 0.001      |         |
|             | 80:MT     | 1.003    | 0.297      |         |
|             | 110:MT    | 1.005    | 0.041      |         |
|             | 80:JP     | -0.999   | 0.229      |         |
|             | 110:JP    | -0.999   | 0.662      |         |
|             | 80:log(Mix)| 1.486 | 0.189      |         |
|             | 110:log(Mix)| -0.934 | 0.831      |         |

| Model       | Random effects standard deviation | Determination coeff. R² |
|-------------|-----------------------------------|-------------------------|
|             | Tract id | Direction | Tree id | Marginal | Conditional |
| Nutrients model | -        | -         | -       | 0.29     | 0.658   |
| Water model  | -        | -         | -       | 0.313    | 0.662   |
| Age model   | -        | -         | -       | 0.25     | 0.644   |
precipitation on BAI in the normal moisture regime W1 (Tab. 2 – W3:JP, p < 0.003). For the other FSS, the effect of precipitation on BAI was positive, and did not differ significantly from the effect of precipitation in normal soil moisture sites W1 (W2:JP, p > 0.18) and especially in peatland soils (W4:JP, p > 0.76).

The results on the “Intercept estimate” for W2, W3 and W4 compared to group W1 when the effect of the other variables in the model were eliminated (equal to 0) shows that BAI in the W1 F1 differed quite significantly from the other groups (Tab. 2 – lines W2-W4, p < 0.01). The highest level of significance was between BAI in the W1 and W4 FSSs (p < 0.05). Thus, the soil moisture regime modifies significantly the effect of the meteorological parameters on pine BAI. The integrated effect of meteorological and stand mixture parameters increased the significance among the BAI in different moisture FSSs, explaining more than 66% of the variation.

Effect of meteorological conditions on pine BAI in different “Nutrients” FSS groups

Both meteorological parameters also had a highly significant positive effect on pine BAI in main mesotrophic (N2) FSS. An increase in March temperature by 1 °C resulted in an increase in pine BAI in all groups of “Nutrients” FSSs, i.e., by 0.24 cm² in N1, 0.19 cm² in N2, 0.20 cm² in N3, and 0.13 cm² in N4 (Fig. S2 in Supplementary material). The only significant difference was found between the effect of temperature on BAI in oligotrophic soils N3 and in mesotrophic soil N2 (Tab. 2 – lines N3-MT, p < 0.001). Quite different effect of temperature on BAI was found in very oligotrophic soils (N4) if compared with data obtained in (N2). Difference was close to the level of significance (N4:MT, p > 0.1). No significant difference was found between the effect of temperature on BAI between eutrophic N1 (N1:MT, p > 0.5) and mesoeutrophic soils N2. This means that March temperature had a similar significant effect on pine BAI at both eutrophic (N1) and mesoeutrophic soils (N2) FSSs.

Similar results were obtained for the effects of June precipitation on BAI at FSSs with different nutrient levels. An increase in precipitation by 10 mm in June resulted in an increase in pine BAI in all groups of “Nutrients” FSSs, i.e., by 0.36 cm² for N1, 0.53 cm² for N2, 0.66 cm² for N3, and 0.03 cm² for N4 (Fig. S2 in Supplementary material). No statistical difference was detected in the effect of precipitation on BAI between eutrophic N1 (N1:JP, p > 0.5) and mesoeutrophic soil N2. The effect of precipitation on BAI of pine trees growing in oligotrophic soils (N3) and very oligotrophic soils (N4) FSSs significantly differed from the effect of precipitation on BAI at mesotrophic (N2) FSSs (Tab. 2 – N3:JP and N4:JP, p < 0.001) and also at eutrophic N1 FSSs.

The results on the “Intercept estimate” for N1, N3 and N4 compared to the FSS N2 value when the effect of the other variables in model were eliminated (equal to 0) revealed that differences in BAI were not statistically significant among “Nutrients model” groups (Tab. 2 – lines N1-N4, p > 0.05). This is in full agreement with the fact that Scots pine has lower demand for nutrients than for water. Further, significant differences in BAI in different nutrient soil conditions were found, mainly depending on age and interactions with meteorological and stand mixture parameters explaining up to 65% in BAI variation. The comparison between “water” and “nutrients” effects revealed that soil moisture was more important for pine BAI than soil nutrients if other parameters are constant.

Effect of tree age on relationships between pine BAI formation and meteorology in different FSS groups

Both meteorological parameters had a significant and positive effect on the BAI of pine trees of ages varying between 40 to 70 years. Tree age had a significant effect on the relationship between temperature and BAI (Fig. S3 in Supplementary material). The relationship between BAI formation and proportion of pine in stands at meso-eutrophic soil and “Nutrients” FSSs, with the exception of the eutrophic N1 FSS, where the effect of stand mixture on BAI was close to the level of significance (p < 0.07). The relationship between BAI formation and proportion of pine in stands at mesotrophic soil (N2) FSSs did not differ significantly from the relationships established in the other “Nutrients” FSS groups, with the exception of the eutrophic N1 FSS, where the effect of stand mixture on BAI was close to the level of significance (p < 0.07).

Effect of deciduous tree proportion in pine stands in different “Water” and “Nutrients” FSSs and age groups

In this study the increase in deciduous tree proportion (DTF) in mixed pine stands (decrease in the “Mix” parameter, i.e., pine proportion in pine stand) also had a positive effect on BAI formation, but only in typical Scots pine FSSs. The negative effect of deciduous trees was found only in forest stands on drained peatland soils W3. For normal moisture FSSs (W1), a 10% reduction in pine proportion (increase in deciduous trees proportion) increased pine BAI by more than 0.5 cm², while in drained peatland (W3) soil FSSs decreased by almost 0.7 cm² (W3:log(mix) – Tab. 2, Fig. S4). The relationships between other FSSs (W2 and W4) showed a similar positive increase in BAI when the pine mixing proportion decreased.

The relationship between BAI formation and proportion of pine in stands at meso-trophic soil (N2) FSSs did not differ significantly from the relationships established in the other “Nutrients” FSS groups, with the exception of the eutrophic N1 FSS, where the effect of stand mixture on BAI was close to the level of significance (p < 0.07). There, a 10% decrease in pine proportion (increase in deciduous tree proportion) re-

---

**Fig. 4** - Comparison of the effect of the 10% and 50% ratio of deciduous trees in pine stands at different FSSs under recent climate changes on annual increments of a single dominating tree in pine stands.
resulted in pine BAI increase by 0.63 cm² in meseuotrophic soil (N2) while in eutrophic soils (N1) in BAI decrease by almost -0.9 cm². At oligotrophic N4, a 10% increase in deciduous trees proportion in pine stands increased pine BAI by 0.2 cm² and 0.07 cm², respectively. This effect of deciduous trees on pine BAI did not differ significantly from that detected in meseuotrophic N2 F2s.

As the F2 fertility decreased from meseuotrophic to very oligotrophic soils, there was a clear decrease in the positive effect of deciduous trees on pine BAI. Pine trees growing in stands with a higher deciduous tree proportion (Fig. S4 in Supplementary material) had a higher BAI than pine trees growing in pure pine stands, especially at N2 F2s. The opposite effect on pine BAI was observed in eutrophic soils (N1). There, deciduous trees inhibited pine BAI formation. This site is not typical for Scots pine in hemi-boreal forests.

Tree age had no significant effect on the relationship between pine BAI and deciduous tree proportion in mixed pine stands. A 10% reduction in pine proportion in a stand (or increase in deciduous tree proportion) resulted in pine BAI increase from approxima- tely 0.9 cm² in younger pine (60) up to 1.1 cm² in older pine (110).

An increase in the proportion of deciduous trees resulted in both increase/de- crease of pine BAI more significantly at FSSs under different moisture regimes vs. different levels of fertility (Fig. 4). A 10% increase in DTP resulted in an increase in pine BAI by approximately 0.9 cm² at W4, a little lower at W2 and the lowest 0.6% at W1, while by only 4.8% at N2, a little lower at N3 and the lowest, only 1.6%, at N4 F5. Age of pine stand had no significant effect on the increase in pine BAI with increasing DTP in mixed stands and made about 7.8% in middle aged and mature stands.

The most significant effect of deciduous trees on pine BAI was observed at N2 and W4 F5s. The positive effect of deciduous trees at N3 and W1 F5s, which dominate in Lithuania, was slightly lower. A negative effect of deciduous trees resulting in the reduction of pine BAI was established only at N1 and W3 F5s, not typical for pine growth in Lithuania.

Fig. S4 (Supplementary material) shows that a DTP up to approximately 50% in meseuotrophic soil N2 F5 resulted in the same pine BAI as in pure pine stands in eu- trophic soils F5 (N1). The basal area increment of pine trees was around 13 cm² per year. On the other hand, DTP of up to 50%, which increased pine BAI in oligotrophic (poor) fertility soils (N3) by up to 9 cm² per year, reduced pine BAI by a similar level in eutrophic soils (N1).

Such opposite effects of the proportion of deciduous tree species, mainly Betula spp., on pine growth should be taken into account in silvicultural measures for sus- tainable development of boreal or hemi- boreal coniferous forest.

Discussion

Main methodological approach of the study

Forest inventory data collected based on rigorous statistical methods provide reliable data on forest productivity since the circle plots are the same size in all sampled areas and are evenly located, forming a tight network in the area of investigation. Caution should be exerted, however, with the use of such datasets to detect the ef- fects of meteorology or new meteorologi- cal data on the productivity of pine forest. Tree ring width data are the most reliable as the accuracy of annual tree ring width data obtained from tree stem analysis is typi- cally higher than DBH data obtained from permanent sample plots of national forest inventories (Garcia 2005). Short time series and poor age records are mentioned among the most concerning problems, to- gether with the fact that dendrometric data are often recorded with much lower precision (Sterba et al. 2014).

Tree rings are also widely used to under- stand the impact of climate change on for- est carbon cycling, and to validate process- based models of forest productivity, where temporal variation of productivity is deter- mined by large trees (Xu et al. 2019). Com- bining conventional forest inventory data and annual increment data based on incre- ment cores of dominant trees in stands of- fers great potential to evaluate growth patterns under different forest conditions in relation to environmental factors. In this study, an evaluation of the capacity of Scots pine to adapt to climate change was undertaken on the basis of Lithuanian Na- tional forest inventory (NFI) data on stand structure and productivity, supplemented with tree ring width series of single domi- nating pine trees. This methodological ap- proach allowed to establish the main ten- dencies in BAI formation of pine trees under different F5 conditions and to detect key meteorological parameters, which af- fect pine forests in Lithuania.

Radial growth of trees within forest stands greatly depends on the interactions between competition and environmental conditions (Piutti & Cescatti 1997). Accord- ing to the current state of knowledge, dominating tree parameters reflect well the productivity of a fully stocked even-aged stand because they are independent of density, or competition intensity (Skovs- gaard & Vanc1ay 2008), and therefore, are used as a measure of site productivity (Monserud & Sterba 1996). Parameters of dominating trees are the key parameters in site index models (Pretzsch et al. 2002; Reed et al. 2003). They are also often in- cluded in forest development models used for the estimation of species mixing effects or of the effect of forestry measures, ex- pressed through variation in stem radial growth (Sterba et al. 2014, Sharma et al. 2019).

It is essential to carefully consider tree sample design in dendrochronological anal- yses, in order to estimate precisely tree in- crement, their changes, and the effects of climate change, minimizing the sampling effort from more than 2000 permanent ob- servation plots of Lithuanian national for- est inventory net. We sampled from 1 up to 2-3 tree cores per plot to ensure reliable representation of mixed pine stand prod- uctivity and productivity variations during the last few decades, even though only one tree ring series per tree species was represented. The results confirmed our hy- pothesis on the relationship between the mean values of pine tree BAI (based on dendrochronological data of dominating single pine trees) and gross stand volume increment (based on NFI data). The Wilcoxon test for all species including Scots pine supported the null hypothesis (p = 0.650- 0.972), meaning that there was no signifi- cant difference in paired series gathered using different methodologies, i.e., between the data collected within the frame of NFI and the data collected from tree- level dendrochronological investigation (Kulbokas et al. 2019). A regression model created with the NFI data on stand volume increment as the dependent variable and tree-level BAI as the independent variable was characterized by significant coeffi- cients of determination exceeding 70%.

To meet the objectives, W1 and N2 F5s were used as references for water and nu- trient levels in forest stands as they were considered to represent best growing con- ditions for Scots pine in Lithuania. Every factor was compared against this refer- ence to evaluate differences in chosen fac- tor groups (soil moisture or nutrition). Such comparisons help to find factors that positively or negatively affect BAI at given soil moisture or nutrient levels.

To reduce BAI variation among the stands, standwise calibration was per- formed based on tree BAI. It is known that BAI captures growth trends better than tree-ring width and avoids the need for data detrending (Blondi & Qeadan 2008). However, in our model, significant relations- hips between BA and BAI remained, indi- cating direct effects of stem diameter and tree age on BAI. This relationship could be attributed to the effect of climate warm- ing, when the observed increase in temper- ature had a more significant effect on the increase in BAI of the older and bigger pine trees than on the younger and thinner trees.

Much of data variability resided in the random effects. Differences between cor- responds to marginal R² (variance explained by fixed factors) and conditional R² (vari- ance explained by fixed and random fac- tors) reflect how much variability is related to random effects (Tab. 2). Unstructured random components (tree and plot loca- tion together with tree id) were included into the models to account for subject-spe- cific BAI variations through the nonlinear mixed-effects modeling approach. In the
present study the introduction of these components helped to explain differences between different growth conditions that could not be accounted for otherwise. Such differences within or between plots of the same soil type may be due factors we did not account for in the model such as genetics or competition. Recently, this approach has been used to increase the prediction accuracy of forest growth models (Fu et al. 2018, Sharma et al. 2019). In such situations it is claimed that a mixed-effects modeling approach is the most appropriate as it takes into account the hierarchical data structure and includes all the subject-specific variations, stochasticity, and randomness present in the data (Chapagain & Sharma 2021). Inclusion of the parameters of dominating trees (BA) and pine stands (mixture, site condition, age) together with subject-specific random effects into the BAI models significantly increased the explanation of BAI variations and the biological importance of the models especially in stands of pine mixture with deciduous tree species, mainly Betula spp.

Effect of deciduous tree proportion (DTP) on pine stem BAI

Forest mixtures offer great potential for silviculture, not only due to the higher resilience and resistance to various biotic and abiotic disturbances but also due to higher productivity and capacity to mitigate risks associated with climate change (Pretzsch et al. 2015, Bouwman et al. 2021). However, their growth dynamics are often difficult to predict because species interactions vary with climatic and edaphic conditions, stand structure and forest management (Aldea et al. 2021).

DTP in pine stands had a beneficial effect that could be explained by better light distribution in the tree canopy and higher nutrient level in the upper soil horizon due to deciduous leaf litterfall. This result agrees well with findings on the beneficial effects of deciduous trees in pine stands, evidenced as increased adaptability of mixed stands to recent changes (Liang et al. 2016), and in some cases, increased productivity when compared to similar mono-species stands (Rio et al. 2016, Jaclet et al. 2018).

Soil properties might affect the interaction between deciduous and pine trees. Jaclet et al. (2018) reported that the relationship between biodiversity and productivity can become more positive or more negative under different soil conditions, especially with limited water or nutrient supply (Ratcliffe et al. 2017, Lackey et al. 2015). As availability of water increases, the availability of nutrients decreases, resulting in an increase in light competition (Pretzsch et al. 2014).

Maestre & Cortina (2004) also claimed that when a limited resource is the only fundamental abiotic stress factor, facilitation could only occur when deciduous trees increased the availability of this resource for the coniferous tree species. Such interactions might possibly be happening in the poorest peatland soils FSs, i.e., at combined W4 and N3 and W4 and N4 FSs, which could explain the increased pine growth at such sites when useful minerals are most likely scarce (oligotrophic FS). These relationships confirm results on very nutrient reach N1 and water regime interrupted drained W3 FS where deciduous trees reduced pine BAI.

The results on the mixture effect in pine forests revealed that the highest magnitude of knowledge that carefully designed DTP on pine stem BAI could bring a wider range of ecosystem services and be more resilient than pure stands, while also increasing productivity (Jaclet et al. 2018, Rio et al. 2016, Pretzsch et al. 2014, 2015).

Effect of meteorology on pine stem BAI

Scots pine growth is affected by frost, heat and drought throughout Europe (Aldea et al. 2021). The radial growth responses of the eastern Baltic population of Scots pine showed explicit regional gradients depending on local climates, indicating gradual shifts in the effects of winter temperature and summer moisture regimes (Mattsons et al. 2021).

The effect of June precipitation on BAI at FSs with different moisture regimes was not as significant as expected. Despite its positive and significant effect on BAI in almost all FSs, a negative effect of precipitation in drained peatland soil (W3) was detected with a slight decrease in BAI. At this FS, a decrease in BAI with increasing proportion of deciduous trees in pine stands was observed, possibly due to light competition with faster growing deciduous trees like aspen, alder or birch. Trees with lower average DBH, but similar height were observed at the drained sites when compared to the other sites except for the peatland sites (W4 – Tab. 1). Similar results were obtained by Hökkä & Groot (2011) with drained sites, as they found that growth might be limited by excess soil water in low stand sites that possibly had damaged drainage systems. An annual report of the Lithuanian Mineral and Mining Sector (Skreinskas et al. 2010) stated that 54% of drained sites are in bad working condition and 15% have insufficient drainage, mainly due to activity of ineffectively managed beaver communities. It is possible that the effect of higher precipitation in beaver-damaged stands resulted in episodes of waterlogged regime causing a stress reaction in pine trees.

On the other hand, changes in tree growth at northern latitudes is mainly limited by the availability of water in sandy soils (Michelot et al. 2012), making tree growth there more sensitive to climatic variation (Lebourgeois et al. 2010). Possibly, the effect of precipitation during June on BAI formation at FSs with lower water availability differed significantly in comparison with the effect observed at waterlogged soil FSs.

Our findings showed that the effect of precipitation in June on pine BAI was more significant than the effect of temperature in March in different nutrient levels, while the effect of temperature in March was more significant than the effect of precipitation in June in FSs with different moisture levels. These findings agreed with our earlier results showing negative effects of drought and positive effects of climate warming on pine basal area increment in locally or regionally polluted forest (Juknys et al. 2002, 2003, 2014, Augustaitis et al. 2015, 2018). No negative effect of increased temperature on pine BAI in the remaining months was established. Therefore increases in early spring temperature and precipitation amount during vegetation could promote Scots pine tree growth and overall yield of pine stands in Lithuania.

Conclusions

From our study we conclude that:
• an increase in mean temperature in March resulted in an increase in dominating pine tree BAI more significant than an increase in precipitation amount in the middle of the vegetation period, when intensity of this increase in older pine stand was higher than in younger ones; it is the key finding of the effect of climate warming on pine BAI in hemi-boreal forests;
• an increase in the proportion of deciduous trees in mixed pine stands had a positive effect on dominating pine tree BAI in typical for Scots pine growth FSs and most importantly in oligotrophic soil (N3) prevailing in Lithuania;
• in drained peatland soils (W3), an increase in precipitation during June and of the deciduous tree ratio in pine stands had a negative effect on dominating pine tree BAI; the same negative tendency was observed in the stands on eutrophic soil (N1);
• tree age had a significant effect only on relationships between dominating pine tree BAI and mean temperature in March; the effect of age on the relationships between pine tree BAI formation and precipitation amount was not significant;
• forestry treatments directed toward the increase of deciduous tree proportion in the most common normal or temporarily waterlogged meso-eutrophic and oligotrophic pine stands might increase not only the biodiversity of pine stands, but also their productivity, and could help the hemi-boreal pine forest adapt to and mitigate the new threats of climate warming.
Acknowledgments

We would like to thank Dr. Peter Biber from the Munich Technical University for helping with the statistical analysis. This research was supported by the Forest Policy Group of the Ministry of Environment of the Republic of Lithuania.

References

Aldea J, Ruiz-Feinado R, Del Río M, Pretzsch H, Heym M, Brazaitis G, Jansons A, Metslaid M, Barbeito I, Bielak K, Granhaus A, Holm SO, Nothdurft A, Sitko R, Lof M (2011). Species stratification and weather conditions drive tree growth in Scots pine and Norway spruce mixed stands along Europe. Forest Ecology and Management 481: 11867-11897. - doi: 10.1016/j.foreco.2011.11.015

Augustaitis A (2007a). Pine sawfly (Diprion pini L.) - Related changes in Scots pine crown defoliation and possibilities of recovery. Polish Journal of Environmental Studies 16: 363-369.

Augustaitis A, Augustaitiene I, Cirga G, Mazeika JuK, Juknys R, Vencloviene J, Augustaitis G, Augustaitis R, Yulis A, Adomas V, Bartkevičius E, Kurkonis N (2014). Dynamic response of tree growth to changing environmental pollution. European Journal of Forest Research 133: 713-724. - doi: 10.1007/s10337-013-0712-3

Kairiukstis L, Grigalas J, Skudlione L, Stravin- skiene V (1987). Physiological and dendrochronological indications of forest decline and their application for monitoring forest decline and development: Regional and Global Consequences” (Kairiukstis L, Nilsson S, Strazak A eds). IASA, Luxemburg, pp. 151-169.

Kulbokas G, Jurevičienė V, Kulesis A, Augustaitis A, Pretuskas E, Mikalajunas M, Vitas A, Mozgeris G (2019). Fluctuations in gross volume increase estimated by the Lithuanian National Forest Inventory compared with annual variations in single tree increment. Baltic Forestry 27: 253-280. - doi: 10.4694/vol27sp2pp273

Kulesis A, Kasperavičius A, Kulbokas G (2016). Lithuania. In: “National Forest Inventories - Assessment of Wood Availability and Use” (Vidal C, Alberdi I, Hernandez L, Redmond J eds). Springer International Publishing, Cham, Switzerland, pp. 521-547. - doi: 10.1007/978-3-319-44515-6_28

Lebourgeois F, Rathgeber CB, Ullrich E (2010). ‘Viability of French temperate coniferous forests to climate variability and extreme events (Abies alba, Picea abies and Pinus sylvestris).’ Journal of Vegetation Science 21: 364-376. - doi: 10.1111/j.1654-030X.2009.00148.x

Liang J, Crowther TW, Picard N, Wiser S, Zhou M, Alberti G, Schulze ED, Mcguire AD, Bozzato F, Pretzsch H, de-Miguel S, Paquette A, Herault B, Scherer-Lorenzen M, Barrett CB, Click HB, Hengeveld GM, Nabuurs GJ, Pfaatsch S, Viana H, Vibrans AC, Ammer C, Schall P, Vertesy D, Tchekobara N, Fischer M, Watson JV, Chen HY, Lei X, Schelhaas MJ, Lu H, Gianelle D, Parfenova E, Salas C, Lee E, Lee B, Kim HS, Bruelheide H, Coomes DA, Plootto D, Sunderland T, Schmid B, Goultier-Fleury S, Sonke B, Tavani R, Zhu J, Brandl S, Vayreda J, Kitahara F, Searle EB, Neldner J, Ngugi MR, Baraloto C, Frizerria L, Ba azy R, Oksýnský J, Zawii A-Nielek wiici B, Bouriaud O, Bussotti F, Finner L, Jaroszwiczki B, Jucker T, Vidal-Salas C, Jagodzinski AM, Peri PL, Bonnade G, Marth W, Obrien T, Martin EH, Marshall AR, Rovero F, Bitarhó R, Niklaus PA, Alvarez-Loayza P, Chamuya N, Valencia R, Mortier F, Wortel V, Engone-Olabio NG, Ferreira LV, Odeke DE, Vasquez RM, Lewis SL, Reich PB (2016). Positive biodiversity-productivity relationship predominant in global forests. Science 354 (6359): 1028-1031. - doi: 10.1126/science.aaf8557

Miacis M, Andreu L, Bosch O, Camarero JJ, Gutierrez E (2006). Increasing aridity is enhancing silver fir Abies alba (Mill.) water stress in its semi-arid steppe. Royal Society Biological Sciences London B277: 5331-5333. - doi: 10.1098/rspb
Seidling W, Ziche D, Beck W (2012). Climate responses and interrelations of stem increment and crown transparency in Norway spruce, Scots pine, and common beech. Forest Ecology and Management 284: 196-204. - doi: 10.1016/j.foreco.2012.07.015

Sharma RP, Stefanick I, Vacek Z, Vacek S (2019). Generalized nonlinear mixed-effects individual tree diameter increment models for beechnets in Slovakia. Forests 10: 451-475. - doi: 10.3390/f10050451

Skovsgaard JP, Vanclyck JK (2008). Forest site productivity: a review of the evolution of dendrometric concepts for even-aged stands. Forestry 81: 12-31. - doi: 10.1093/trees/ftp041

Skrinskas S, Gasiuniene VE, Laurinavius A, Podagelis I (2010). Lithuanian mineral resources: their reserves and possibilities for their usage in road building. The Baltic Journal of Road and Bridge Engineering 5 (4): 218-228. - doi: 10.3846/bjee.2010.30

Schiepker H (1999). Overview of recent growth trends in European forests. Water, Air, and Soil Pollution 116 (1/3): 33-46. - doi: 10.1023/A:1005225195952

Sterba H, Del Rio M, Brunner A, Condes S (2014). Effect of species proportion definition on the evaluation of growth in pure vs. mixed stands. Forest Systems 23: 547-559. - doi: 10.14245/201423306051

Tocchi EC (2017). Forest and environment: developments in global change ecology. Forestry Research and Engineering 1 (3): 100-105. [online] URL: http://www.researchgate.net/publication/325627538

Xu K, Wang X, Liang P, Wu Y, An H, Sun H, Wu P, Wu X, Li Q, Guo X, Wen X, Han W, Liu Ch Fan D (2019). A new tree-ring sampling method to estimate forest productivity and its temporal variation accurately in natural forests. Forest Ecology and Management 433: 219-227. - doi: 10.1016/j.foreco.2018.10.066

Supplementary Material

Fig. S1 - Effect of the mean temperature in March and precipitation amount during June on annual BAi at FSs with different moisture regimes, i.e., “Water” FS group.

Fig. S2 - Effect of mean temperature in March and precipitation amount during June on the annual BAi for the considered 20-year period at FSs with different nutrient levels, i.e., “Nutrients” FS group.

Fig. S3 - Effect of temperature in March and precipitation during June on BAi formation of pine trees in different aged FS groups.

Fig. S4 - Effect of the proportion of pine trees in mixed pine stands on pine BAi formation at FSs with different moisture regimes, “Water”, and fertility levels, “Nutrients”.

Link: Augustaieni.e_3703@suppl001.pdf