How Covid-19 Pandemic Influenced Air Quality in Polish Cities – Lessons from Three Lockdowns

Katarzyna Lindner-Cendrowska¹ • Kamil Leziak² • Peter Bröde³

¹ Institute of Geography and Spatial Organization
Polish Academy of Sciences
Twarda 51/55, 00-818 Warsaw: Poland
e-mail: klindner@twarda.pan.pl (corresponding author)

² Faculty of Geography and Regional Studies
University of Warsaw
Krakowskie Przedmieście 30, 00-927 Warsaw: Poland
e-mail: k.leziak@uw.edu.pl

³ Leibniz Research Centre for Working Environment and Human Factors
at TU Dortmund (IfADo)
Ardeystrasse 67, 44139 Dortmund: Germany
e-mail: broede@ifado.de

Abstract
The aim of this study was to determine how COVID-19 pandemic influenced air quality in the chosen Polish cities. Data on nitrogen oxides, carbon monoxides, fine and coarse particulate matter concentrations from air quality monitoring stations was used to compare pollutants levels during the pandemic and in the 5-year pre-pandemic period. The impact of the pandemic on the air quality has been analysed using linear mixed effect models, adjusting for long-term, seasonal and weekly trends and meteorological conditions. Results showed that during the pandemic, until the second lockdown only nitrogen oxides levels were significantly reduced (up to 20%), while when again loosening restrictions the rebound effect led to 20-30% increase of all analysed pollutants.

Key words
air pollution • COVID-19 • lockdown • nitrogen oxides • carbon monoxide • particulate matter

Introduction
The COVID-19 pandemic has affected lives of many people all over the world. First cases of a new disease were reported in Wuhan, Hubei Province of China on December 12, 2019. On January 7, 2020, a previously-unknown coronavirus had been identified as an agent for the ongoing outbreak of the disease. In following weeks first imported cases of the new coronavirus disease had been detected in numerous countries, including
Thailand, Japan, Germany, Canada and USA. Following a rapid global outburst of infection in February and March, on March 11 COVID-19 had been declared a pandemic by World Health Organisation (CDC, 2021).

The first case of COVID-19 in Poland had been officially confirmed on March 4, 2020 (Ministry of Health, 2020). For the first week afterwards daily number of new cases stayed below 10. Polish government quickly undertook legislative steps to reduce possible spread of the virus, introducing the first lockdown from March 12. It included the closure of all educational institutions, as well as numerous cultural institutions like theatres, cinemas, libraries and museums. It was also strongly recommended to introduce remote work wherever possible. Despite government’s efforts daily number of new cases continued to grow, reaching triple-digit value on March 21, 2020. Additional restrictions had been implemented from April 1, 2020, including limited number of customers in shops, closure of the most of service points, parks and recreational areas, along with the curfew for underage (the ban for people under the age of 18 to stay in public space without the company of an adult). Some of the restrictions were lifted at the end of April and most of them on June 6, 2020. Because of the introduced lockdown, from April till late July the daily number of new cases stayed below 500, but then began to rise, peaking at 903 new cases on August 21, 2020. That Spring and Summer spread of coronavirus is often marked as the so-called “first wave” of the pandemic in Poland (Chief Sanitary Inspectorate, 2021).

Due to spiking number of new cases (over 10 thousand on October 21, 2020 and over 20 thousand new cases on October 29, 2020), the government decided to introduce a second lockdown on November 9, 2020. It was preceded by a maximum of 27,875 new cases of COVID-19 on November 7, indicating a peak of the so-called “second wave”. The second lockdown resulted in closure of all elementary and secondary schools, cultural institutions, limited number of people allowed in churches and shops as well as shopping malls closure (which were later temporarily reopened in December). The lockdown proved to be successful, with daily number of new cases gradually dropping to 2-7 thousand in the second half of January 2021. Shopping malls and some cultural institutions were fully reopened on February 1, 2021, although some restrictions still applied (i.e. bars and restaurants only allowing take-out orders).

Daily number of new cases of COVID-19 started to rise again abruptly in the second half of February 2021, foreshadowing the so-called “third wave” of the pandemic. It had two distinctive peaks of over 35 thousand cases, on March 26 and April 1, 2021. On March 27 the government introduced a third lockdown, again closing shopping malls and limiting number of people allowed in shops, churches and other institutions, as well as forbidding public gatherings of more than five people. Nurseries and kindergartens were closed for general public (only medical and public service staff’s children were admitted) and schools of all levels provided only remote teaching. Restrictions were being lifted gradually from April 19, 2021 to May 29, 2021.

The COVID-19 pandemic and associated lockdowns resulted in changes in people’s mobility, affecting it in multiple ways (Parr et al., 2020). Due to sudden closure of schools at the beginning of pandemic many parents had to stay at home with their children, as they were unable to find a caregiver. Some of workplaces had been closed initially, but later many companies adapted to remote work, as lockdowns progressed and further restrictions have been introduced (i.e. a so-called “national quarantine” and stay-at-home orders). Additionally, the government introduced ad hoc legislation allowing for some public services and administrative procedures to be undertaken remotely, further reducing the necessity to leave homes. Closure of shopping malls and other points of interest also resulted in more people staying at home. Since daily commute is one of main sources of traffic (USDT, 2010), changes in people’s mobility could have affected concentration
of certain air pollutants produced by Internal Combustion Engine vehicles.

Despite the relatively short time span of the COVID-19 pandemic, many researches have already been conducted to determine the impact of this event on air pollution and air quality. Various approaches were applied, depending on the availability of the data. In the majority of cases the atmospheric concentration of NO₂ has been used as the main indicator of air pollution, for its strong correlation with traffic intensity and for being one of the most commonly measured pollutants (Marinello et al., 2021). Many researches present pandemic-related changes in concentration of both gaseous emissions like NO₂ and SO₂ as well as particulate matter like PM10 or PM2.5 (e.g. Berman & Ebisu, 2020; Kerimray et al., 2020; Silver et al., 2020; Adam et al., 2021; Gao et al., 2021). In some works air quality was inferred from remote sensing data, i.e. satellite imagery products from MODIS or TROPOMI (Filanchyk et al., 2021a, 2021b; Hammer et al., 2021; Venter et al., 2020; Wyche et al., 2021).

Numerous studies show that lockdowns caused air pollution to decrease (Brimblecombe & Lai, 2020; Dong et al., 2021; Dutheil et al., 2020; Liu et al., 2021; Rodríguez-Urrego & Rodríguez-Urrego, 2020; Rossi et al., 2020; Skirienė & Stasiškienė, 2021; Varga-Balogh et al., 2021). However, many of them have only considered the first lockdown, happening between March and May 2020 in most countries (e.g. Baldasano, 2020; Cameletti, 2020; Collivignarelli et al., 2020; Piccoli et al., 2020; Sharma et al., 2020; Singh & Chauhan, 2020; Moreda-Piñeiro et al., 2021). Therefore, it is important to determine whether subsequent lockdowns introduced as a response to “waves” of the pandemic had a similar effect. To our knowledge, so far there has been no studies providing a complex analysis of how multiple lockdowns impacted air pollution in European cities during the COVID-19 pandemic.

In virtually all researches, the reduced concentration of air pollutants during lockdowns has been determined through a comparison of air quality and pollution data from lockdowns and a pre-pandemic reference period. In many cases only the single year preceding the COVID-19 outbreak (2019) has been used as a reference (Chen et al., 2021; Dantas et al., 2020; Kumari & Toshniwal, 2020; Parr et al., 2020; You et al., 2021). In others a reference period has been defined through averaging data from a few consecutive years prior to the pandemic (Chen et al., 2020; Connerton et al., 2020; Ginzburg et al., 2020; Kerimray et al., 2020; Zangari et al., 2020).

Typically, the concentrations of air pollutants, especially particulate matter, show a distinct seasonal pattern in Polish cities. In winter, air temperature drops are accompanied by increases in particulate matter levels, related to the rising heating intensity (Nidzgorska-Lencewicz & Czarnecka, 2015). Various studies show that the weather affects air quality and the concentration of pollutants (Nidzgorska-Lencewicz & Czarnecka, 2015; Cichowicz et al., 2017; Jędruszkiewicz et al., 2017; Kamińska, 2019; Xu et al., 2019; Kalbarczyk & Kalbarczyk, 2020; Zhu et al., 2021), and deterioration of air quality is usually associated with the occurrence of unfavourable weather conditions (Reizer & Juda-Rezler, 2016). Although level of nitrogen oxides is strongly positively correlated with the intensity of road traffic, it is generally most dependent on meteorological factors (Cichowicz et al., 2017), among which the increase in wind speed, that stimulates mixing processes and enables dispersion of air pollution, proved to be the most important (Kamińska, 2019; Kalbarczyk & Kalbarczyk, 2020). The air quality can also improve under wet deposition conditions, as suggested by high negative correlations between rainfall and air pollution found in Polish cities (Jędruszkiewicz et al., 2017).

Yet there have only been few researches so far that tried to evaluate the influence of meteorological conditions during a lockdown on air pollution, in addition to the impact of pandemic-related mobility restrictions themselves (Bolaño-Ortiz et al., 2020; Fu et al., 2020; Dabbour et al., 2021; Gao et al., 2021;
Gkatzelis et al., 2021). Some studies focus on how certain meteorological elements (i.e. the air temperature, downward UV radiation, humidity) or phenomena (i.e. thermal inversion) can influence the number of COVID-19 cases (Higham et al., 2021; Keikhosravi & Fadavi, 2021; Werner et al., 2021). There has also been an attempt to use a chemistry transport model to determine the impact of the lockdown on the atmosphere’s composition, while offsetting meteorological conditions (Menut et al., 2020). A different approach applied machine learning tools to offset meteorological conditions during the lockdown in Wuhan (Cole et al., 2020). In another study, artificial neural networks were used to model possible air quality improvement depending on lockdowns’ severity and length (Tadano et al., 2021).

It should be noted that not in all cases introducing mobility restrictions and “stay at home” order resulted in an improvement of the air quality. Some studies showed no significant change in the average concentration of air pollutants like NO$_2$, ozone or PM2.5 (Jia et al., 2020; Sicard et al., 2020), while others observed increased concentration of SO$_2$ during lockdown, mostly attributed to changes in power generation and not to traffic-related sources (El-Sayed et al., 2021). There have been cases of increased ozone pollution due to reductions in NO emission caused by the lockdown (Briz-Redón et al., 2021; Jephcote et al., 2021; Ropkins & Tate, 2021). In some cases, decreased concentrations of pollutants during lockdowns have been overwhelmed by unfavourable weather conditions that still caused severe air pollution events, even while the traffic was significantly reduced (Wang et al., 2020).

Few papers show the impact of COVID-19 mobility restrictions on air quality in Poland. They focus on the first lockdown only and use either satellite data (Filonchyk et al., 2021a, 2021b; Grzybowski et al., 2021) or ground data from air quality monitoring stations (Rogulski & Badyda, 2021) and mobile measurements (Polednik, 2021) to investigate changes in air pollutants concentration during one particular period of time. Therefore, it is crucial to check what changes in air quality have occurred as the COVID-19 pandemic unfolded and verify whether introduced administrative regulations had the same impact on air pollution during three subsequent lockdowns in Poland.

The aim of this study was to determine how the concentration of certain air pollutants, i.e. nitrogen oxides (NO$_x$), carbon monoxides (CO), coarse (PM10) and fine particulate matter (PM2.5), in the selected 12 Polish cities has changed during multiple COVID-19 lockdowns, including imposed mobility and service activities restrictions. Consecutive phases of the pandemic were compared to the corresponding periods in the preceding 5 years (2015-2019), taking into account the long-term, seasonal and weekly trends of particular pollution concentrations. Moreover, to account for the possible confounding factors, mean daily air temperature and wind speed, as well as the type of monitoring station surroundings were included in the analysis as covariates.

Materials and methods

Study period

The study period covered one “pre-pandemic” and seven “pandemic” periods from January 1, 2015 to April 18, 2021. The first coronavirus infection in Poland was confirmed on March 4, 2020. Since March 12, the government announced an epidemic emergency state and with the closure of educational and cultural institutions the first lockdown began (Website of the Republic of Poland, 2021). Table 1 presents the 8 phases together with a description of the preventive measures and administrative restrictions introduced in response to changing COVID-19 incidence and death rates. In the analysed period, lockdowns were introduced three times, with a complete shift to distance learning in schools and universities, the closure of cultural institutions, sports and large commercial facilities. The limited mobility of citizens lead to significant traffic reduction (Warsaw Municipal Roads Management, 2021).
The analysis covered 12 out of the 15 most populated cities in Poland, where complete measurement series of at least 3 out of 4 analysed pollutants (NOx, CO, PM10 and PM2.5) were recorded at one spot over a long enough period of time since 2015. All studied towns are important cultural and administrative centres in their region, providing numerous jobs and education at all levels. Each of the cities, just before the outbreak of the pandemic, had more than 200,000 inhabitants and its population density ranged from 3469 pers./km² in Warszawa to 1362 pers./km² in Częstochowa (Tab. 2) (Statistics Poland, 2021).

Air pollution data

In order to characterize aerosanitary conditions in Polish agglomerations in the studied period (01.01.2015-18.04.2021), hourly means of NOx, CO, PM2.5 and PM10 concentration were obtained from 12 air quality monitoring stations of the Chief Inspectorate of Environmental Protection (GIOŚ), located within the borders of the selected cities. Each city was represented by one measuring spot in order to ensure equal representativeness of individual cities in the model. Two types of automatic urban stations were included in the study: 6 traffic stations – located in close proximity to busy roads and monitoring the level of pollution caused mainly by traffic...
Table 2. Population of the studied cities; as of January 1, 2021 (Statistics Poland 2021) and air quality monitoring stations characteristics

| City name and acronym | Population | Population density per km² | Polish Air Quality System identification number | Geographical location (WGS84) | Station type | Measured pollutants |
|-----------------------|------------|----------------------------|-----------------------------------------------|------------------------------|-------------|--------------------|
| Bydgoszcz (BDG)       | 346,739    | 1,955                      | KpBydPlPozna                                   | Φ 53.121774 λ 17.987883      | traffic     | NOx, CO, PM10, PM2.5 |
| Białystok (BIA)       | 297,585    | 2,908                      | PdBialWaszyn                                   | Φ 53.126689 λ 23.155869      | background  | NOx, CO, PM2.5     |
| Częstochowa (CZE)     | 219,278    | 1,362                      | SiCzestoArmK                                   | Φ 50.817217 λ 19.118997      | traffic     | NOx, CO, PM10      |
| Gdańsk (GDA)          | 471,525    | 1,797                      | PmGdaLeczko08                                  | Φ 54.380279 λ 18.620274      | background  | NOx, CO, PM10, PM2.5 |
| Gdynia (GDY)          | 245,867    | 1,813                      | PmGdyPoreb04                                   | Φ 54.560836 λ 18.493331      | background  | NOx, CO, PM10      |
| Kraków (KRA)          | 780,981    | 2,386                      | MpKراكAIKras                                   | Φ 50.057678 λ 19.926189      | traffic     | NOx, CO, PM10, PM2.5 |
| Łódź (LDZ)            | 703,186    | 2,390                      | LdLodzGdansk                                   | Φ 51.775411 λ 19.4509        | background  | NOx, CO, PM10      |
| Lublin (LUB)          | 339,547    | 2,296                      | LbLubObywate                                   | Φ 51.259431 λ 22.569133      | background  | NOx, CO, PM10, PM2.5 |
| Radom (RAD)           | 210,532    | 1,872                      | MzRadTochter                                   | Φ 51.399084 λ 21.147474      | background  | NOx, CO, PM10, PM2.5 |
| Toruń (TOR)           | 201,106    | 1,716                      | KpToruKaszow                                   | Φ 50.017628 λ 18.612808      | traffic     | NOx, CO, PM10      |
| Warszawa (WAW)        | 1,793,579  | 3,469                      | MzWarAlNiepo                                   | Φ 52.219298 λ 21.004724      | traffic     | NOx, CO, PM10, PM2.5 |
| Wrocław (WRO)         | 643,782    | 2,192                      | DsWrocAlWisn                                   | Φ 51.086225 λ 17.012689      | traffic     | NOx, CO, PM2.5     |

and 6 background stations – located within residential areas, characterizing emissions caused primarily by households. Industrial monitoring stations were excluded from the analysis, as their number in the biggest cities in Poland is too small. A full list of monitoring points, including their geographical location, Polish Air Quality System identification number and characteristic is shown in Table 2.

**Meteorological data**

Weather conditions that might have affected the concentration of air pollutants were described in this project by the air temperature (°C, 2 m AGL), wind speed (m/s, 10 m AGL) and total precipitation (mm). All weather data have been acquired from the ERA5-Land reanalysis (Muñoz Sabater, 2019) with an hourly resolution, from January 1, 2015 (beginning of the comparison period) to April 19, 2021 (the end of the last lockdown described in this study). The closest grid points of the reanalysis, corresponding to the air quality monitoring stations’ location, were used. Meteorological data have been processed in R software. Air temperature was converted from Kelvins to degrees Celsius, total precipitation was converted from meters to millimetres of water column, while wind speed was calculated from eastward (U) and northward (V) wind components following the formula (ECMWF, 2021):

$$|\mathbf{V}| = \sqrt{U^2 + V^2}$$
Data analysis and statistics

From the hourly time series data, daily mean and maximum values were calculated for the pollutants (NO\textsubscript{x}, CO, PM2.5, PM10) as dependent variables for subsequent analyses. As predictors, daily mean hourly air temperatures and wind speed were computed as well. Dichotomizing precipitation, we included a binary predictor scoring a day with precipitation, if the total daily precipitation exceeded 1 mm. For the long-term trend assessment, we counted the days from the beginning of the measurement period with day 1 referring to 01.01.2015, which were then converted into years by dividing the day number by 365.25. Seasonal and weekly trends were assessed by month of the year and day of the week, respectively. In addition, days which were neither national holidays nor weekends were classified as ‘labour days’. The city and type of the weather station (traffic or background) were considered as factors. Finally, days were categorized according to the date into the phases defined in Tab. 1, contrasting the pre-pandemic period (phase 1) with the different phases (2-8) of the pandemic.

As the daily mean and maximum values of the pollutants were skewed, the log-transformed dependent variables were analysed for the effect of the pandemic phases in contrast to the pre-pandemic period by linear mixed effects models (Pinheiro & Bates, 2000) with random intercept and slope of long-term trend for city, while adjusting for seasonal (month) and weekly (day of week) cycles, for the effects of the monitoring station type, of labour days, days with precipitation, and daily mean wind speed and air temperature. A first-order autoregressive error structure was assumed for considering the serial correlation within the time series.

The mixed models were fitted with the R statistical computing environment (R Core Team, 2021) using the package nlme (Pinheiro et al., 2021). The effects of the pandemic phases on pollutant concentrations compared to the pre-pandemic period were obtained from the fitted mixed models using the package emmeans (Lenth, 2021), which delivered the corresponding contrasts with 95% confidence intervals (CI) adjusted for multiple comparisons by the Dunnett procedure. The resulting differences were expressed as percentage change computed as \% change = (exp(x-1)\times100, with x denoting the values of the estimated contrasts for the log-scaled dependent variables and their adjusted 95% CI, respectively.

Another set of mixed models was fitted additionally including the interaction of the phases with the type of the weather station allowing for calculating the contrasts (% change) stratified for the station type (traffic or background).

Results

The concentrations of analysed pollutants, especially particulate matter, showed a clear seasonal pattern in Polish cities, as exemplified for the data from Bydgoszcz (BDG) and Lublin (LUB) in Figure 1. During the COVID-19 pandemic in Poland, this seasonal trend still prevailed in case of PM2.5 and PM10 pollution at most analysed monitoring stations, but was less pronounced for NO\textsubscript{x} and CO concentrations exhibiting more even distributions over the year (Fig. 2). The NO\textsubscript{x} pollution was also characterized by the largest range of variation among all analysed pollutants.

During the pandemic in Polish cities (phases 2-8), mean hourly concentrations of all analysed pollutants were lower than in 5-year pre-pandemic period (after Tab. 3). The most significant decrease was observed for NO\textsubscript{x} on traffic stations, where mean hourly values were on average 24.3% lower than in the pre-pandemic period. In the two largest cities, Warszawa (WAW) and Kraków (KRA), the average decrease in hourly NO\textsubscript{x} concentrations from pre-pandemic levels was 35% and 28%, respectively. Interestingly, the opposite direction of changes in pollutant concentrations was observed in Gdańsk (GDA) and Gdynia (GDY), where PM10 levels where on average 10% and 24% higher, than in pre-pandemic period.
Figure 1. Daily mean and maximum hourly pollutant concentrations measured between 01.01.2015 and 18.04.2021 in Bydgoszcz (BDG, left panel) and Lublin (LUB, right panel) as examples for traffic and background monitoring stations, respectively, with the vertical reference lines denoting the transition from the pre-pandemic to pandemic phases.

Geographia Polonica 2022, 95, 3, pp. 255-274
Figure 2. Daily mean and maximum hourly concentrations of the pollutants NO$_x$ (A), CO (B), PM2.5 (C) and PM10 (D) measured during the pandemic period between 12.03.2020 and 18.04.2021 at background and traffic monitoring stations in twelve Polish cities (with abbreviations as in Table 2). Grey shaded periods indicate even pandemic phase numbers (2, 4, 6 and 8 in light grey) and odd numbers (3, 5, 7 in dark grey), respectively. Time range of the phases as in Table 1.
During the COVID-19 pandemic period covered by this analysis, both traffic and background stations recorded lower average air temperature (on average 0.8°C and 0.6°C respectively) and slightly lower wind speed than in the 5-year pre-pandemic period (Tab. 4). This may indicate that meteorological conditions during the pandemic (phases 2-8) were on average more conducive to the accumulation of pollutants than in the reference pre-pandemic period, which could have an impact on the outcome and therefore was consequently adjusted for by the statistical modelling.

Figure 3 summarizes the percentage change in daily mean (Fig. 3A) and maximum

### Table 3. Means (SD) of daily mean hourly air pollutant concentrations measured during pre-pandemic and pandemic periods summarized for all stations as well as stratified for station type and city, respectively. (NA denotes missing statistics)

|                | NOx [µg/m³] | CO [mg/m³] | PM2.5 [µg/m³] | PM10 [µg/m³] |
|----------------|-------------|------------|---------------|--------------|
|                | pre- pandemic | pre- pandemic | pre- pandemic | pre- pandemic |
| Overall        | 69 (71)     | 54 (54)    | 0.48 (0.27)   | 0.45 (0.22)  |
|                | 24 (19)     | 20 (14)    | 34 (25)       | 30 (19)      |
| Station type   |             |            |               |              |
| Background     | 27 (21)     | 22 (16)    | 0.39 (0.19)   | 0.38 (0.18)  |
|                | 21 (17)     | 19 (14)    | 28 (20)       | 26 (16)      |
| Traffic        | 111 (79)    | 84 (60)    | 0.58 (0.31)   | 0.52 (0.24)  |
|                | 28 (22)     | 21 (14)    | 40 (29)       | 34 (21)      |
| City           |             |            |               |              |
| BDG            | 63 (34)     | 51 (38)    | 0.49 (0.24)   | 0.44 (0.23)  |
|                | 23 (19)     | 17 (13)    | 36 (24)       | 29 (22)      |
| BIA            | 18 (12)     | 16 (12)    | 0.37 (0.12)   | 0.34 (0.13)  |
|                | 17 (13)     | 15 (12)    | NA            | NA           |
| CZE            | 92 (62)     | 81 (43)    | 0.53 (0.33)   | 0.52 (0.23)  |
|                | NA          | NA         | 40 (31)       | 36 (20)      |
| GDA            | 25 (21)     | 21 (18)    | 0.37 (0.14)   | 0.34 (0.11)  |
|                | 13 (9)      | 17 (13)    | 20 (12)       | 22 (16)      |
| GDY            | 17 (13)     | 14 (11)    | 0.32 (0.16)   | 0.32 (0.14)  |
|                | NA          | NA         | 17 (11)       | 21 (15)      |
| KRA            | 197 (79)    | 142 (69)   | 0.85 (0.39)   | 0.69 (0.27)  |
|                | 38 (28)     | 25 (15)    | 57 (38)       | 41 (23)      |
| LDZ            | 34 (22)     | 29 (17)    | 0.51 (0.21)   | 0.49 (0.17)  |
|                | 25 (19)     | 17 (13)    | 25 (17)       | 30 (17)      |
| LUB            | 34 (26)     | 25 (15)    | 0.36 (0.21)   | 0.39 (0.21)  |
|                | 24 (18)     | 21 (16)    | 32 (21)       | 25 (17)      |
| RAD            | 33 (21)     | 29 (15)    | 0.43 (0.20)   | 0.41 (0.20)  |
|                | 25 (18)     | 21 (13)    | 34 (21)       | 29 (16)      |
| TOR            | 34 (24)     | 28 (22)    | 0.36 (0.15)   | 0.31 (0.14)  |
|                | NA          | NA         | 26 (16)       | 26 (17)      |
| WAW            | 139 (71)    | 90 (45)    | 0.59 (0.20)   | 0.54 (0.17)  |
|                | 25 (15)     | 20 (13)    | 41 (20)       | 39 (20)      |
| WRO            | 132 (58)    | 112 (54)   | 0.63 (0.23)   | 0.61 (0.21)  |
|                | 24 (19)     | 20 (13)    | NA            | NA           |

### Table 4. Daily mean (SD) of hourly values of air temperature and wind speed, respectively, and of daily total precipitation measured during pre-pandemic and pandemic periods summarized for all stations as well as stratified for station type

|                | Air temperature [°C] | Wind speed [m/s] | Total precipitation [mm] |
|----------------|----------------------|------------------|-------------------------|
|                | pre- pandemic        | pre- pandemic    | pre- pandemic           |
| Overall        | 9.5 (8.4)            | 3.3 (1.3)        | 2.1 (3.7)               |
|                | 8.7 (8.2)            | 3.1 (1.1)        | 2.2 (4.1)               |
| Station type   |                      |                  |                          |
| Background     | 9.1 (8.5)            | 3.2 (1.3)        | 2.2 (3.8)               |
|                | 8.5 (8.2)            | 3.0 (1.1)        | 2.4 (4.4)               |
| Traffic        | 9.8 (8.2)            | 3.4 (1.3)        | 1.9 (3.6)               |
|                | 9.0 (8.1)            | 3.2 (1.1)        | 2.0 (3.8)               |
Figure 3. Estimated percentage change of daily A) mean and B) maximum hourly concentrations of the pollutants NOx, CO, PM2.5 and PM10 in the pandemic phases 2-8 (Tab. 1) compared to the pre-pandemic period for the overall effect from all stations and stratified for station type (traffic, background). Error bars indicate the 95% CI adjusted for multiple comparisons by the Dunnett procedure, with open symbols marking statistically non-significant (ns) effects and filled symbols indicating statistically significant effects, respectively.
(Fig. 3B) hourly pollutant concentrations displayed as overall effect and separately for traffic and background stations, respectively. Figure 3A revealed significant reductions of the daily mean hourly values during the pandemic restrictions for NOx only, with peak reductions of about 20% occurring during the two nationwide lockdown periods (phase 2 & 6) and the preparatory phase 5 preceding the second lockdown. No significant reductions were obtained for the other pollutants except of a reduction of PM2.5 concentration on traffic stations in phases 4 & 5. On the other hand, in phase 7, characterized by loosening the mobility restrictions, a general strong rebound effect with significant increases of more than 30% for PM10 and about 20% for the other pollutants was observed, returning to pre-pandemic values in the subsequent third lockdown period (phase 8).

The corresponding effects stratified for the station type, showed slightly stronger and statistically significant NOx reductions at traffic monitoring points. During the initial pandemic phases, traffic stations reacted earlier by reducing the average daily pollution level, but in phase 7 they showed similar patterns for the rebound effect as background stations (Fig. 3A).

The analogous results for the daily maximum hourly concentrations (Fig. 3B) largely corroborated the results observed for the daily means, though the effects were less pronounced and occasionally differed from the results for the means, e.g. by showing an initial increase during phase 2 of PM2.5 and CO at the background stations and for PM10 concentrations overall.

**Discussion and conclusions**

The implementation of social distancing measures to control the spread of the COVID-19 pandemic had serious consequences, including changes in air quality in different countries. Drastic and sudden restrictions on the mobility and activity of millions of people resulted in significant changes in anthropogenic emissions of air pollutants, which created an unprecedented opportunity to analyse the environmental effects of this real-world experiment (Gkatzelis et al., 2021; Onyeaka et al., 2021). The impact of the first shutdown in Spring 2020 on air quality improvement has been already demonstrated in many studies (Adam et al., 2021; Marinello et al., 2021; Skiriene & Stasiškiene, 2021). However, most of the papers published in the first months after the outbreak of the pandemic, when comparing data from the lockdown period with earlier reference observations, ignore temporal variability and long-term trends in pollutant concentrations, or do not incorporate a specific method to quantify the impact of meteorological conditions on air pollution, which could distort the interpretation of the results (Zangari et al., 2020; Gkatzelis et al., 2021).

As a consequence of stay-at-home orders and decrease in vehicle usage, especially traffic emission was limited, what resulted in much lower NOx and NOx concentrations recorded all over the world (Venter et al., 2020; Briz-Redón et al., 2021; Dong et al., 2021; El-Sayed et al., 2021; Ropkins & Tate, 2021; Varga-Balogh et al., 2021; Wyche et al., 2021). However, the air quality improvement effect was not seen everywhere and for all pollutants. In Hungary, although NOx concentrations dropped up to 45%, the PM10 level even slightly increased (+7%) at the beginning of the first lockdown and was rather weather-driven (Varga-Balogh et al., 2021). Chen et al. (2020) noticed that both NO2 and CO levels were dropping in US cities (up to 49% and 37% respectively) and tended to increase with local population density, while significant reductions of particulate matter occurred seldom and only in those cities with highest NO2 decline. Furthermore, increases in SO2 concentrations in the central and southern regions of Florida were associated with higher power generation due to increased residential usage during lockdown (El-Sayed et al., 2021), while in NYC, in comparison to the same span of time in years 2015-2019 no or minimal improvements of air quality were observed (Zangari et al., 2020).
In turn, Wyche et al. (2021) observed spikes in PM2.5 and PM10 levels alongside lockdown in March 2020 in the south-eastern regions of the UK, but these were probably caused by interregional pollution transport rather than governmental restrictions.

The impact of the restrictions introduced in Poland in response to the development of COVID-19 pandemic on air quality is ambiguous and has not been clearly defined so far. In this study we demonstrated that in general, up to the second lockdown (November 2020 – January 2021), the pandemic significantly reduced NO\textsubscript{x} levels in large Polish cities by 6% to 21% in daily mean and by 3% to 19% in daily maximum concentrations compared to the pre-pandemic period. Similar decrease of NO\textsubscript{2} pollution on urban air quality stations in Poland during first lockdown (April 2020 (~20%)) was observed by Grzybowski (2021), while slightly weaker effect (6-11%) was found for this period by Rogulski and Badyda (2021), which, in comparison to other European cities, represents a relatively low reduction (Solberg et al., 2021). More significant drops in NO\textsubscript{x} concentration were recorded at traffic monitoring stations than at background stations, which has been also confirmed by the results from UK (Jephcote et al., 2021; Ropkins & Tate, 2021) and may be associated with a significant decrease in car traffic in Polish cities in the first stages of the pandemic (Warsaw Municipal Roads Management, 2021). In the case of particulate matter pollution, we observed that the overall effect of pandemic restrictions on pollution level was insignificant in the first 6 phases of the pandemic. However, during the first lockdown, we observed statistically significant increase of the daily maximum hourly values of PM2.5 and PM10 by 13% and 16%, respectively. For the same period, Polednik (2021), as well as Rogulski & Badyda (2021), obtained consistent results from stationary measurements in Polish cities, while Grzybowski et al. (2021), on the contrary, noted significantly lower (~20%) PM2.5 concentration as compared to the 10-year average. Nevertheless, the authors of the latter work drew attention to the fact that the COVID-19 lockdown was not the unique reason for the aerosol concentrations drop during spring 2020, as the meteorological conditions in this period were particularly conducive to good air quality (i.e. lower emissions and dispersion of pollutants). Therefore in our study, to calculate the impact of the pandemic restrictions on pollutants concentrations, we accounted for confounding effects of meteorological conditions and temporal trends, which could have contributed to the fact that we did not observe a reduction effect on PM2.5 and PM10 levels during the first lockdown. On the other hand, conforming to observations in China (Gao et al., 2021), the concentration of particulate matter in Poland, especially PM2.5, started to decrease later. However, this effect was statistically significant only at traffic stations in phase 4 and 5 of the pandemic, which could have been the result of a gradual adaptation of residents to the situation – switching to remote work and distant learning, resulting in less presence of people within the city.

From the beginning of 2021, pollution reductions in relation to the pre-pandemic period were no longer registered. Moreover, in phase 7, statistically significant increases (20-30%) of all analysed pollutants levels (NO\textsubscript{x}, CO, PM2.5 and PM10) were observed. This rebound effect could be related to the increase in the mobility of people who, after a long period of the restrictions, were tired of isolation and limited access to various types of services; hence, they began to move around the city more frequently. At the same time, the increase in pollution concentrations could alternatively be related to shift in travel mode away from public transport towards private cars as a reaction to the threat of coronavirus infection, as was observed during working days in Taiwan (Chang et al., 2021). In addition, in January 2021, the population vaccination program against COVID-19 was launched, which could have contributed to a reduction in the sense of danger in the society and an increase in social and economic activity. However, such high levels of pollution in phase 7 could be related primarily
with adverse weather conditions, which can contribute to the air quality significantly and may outweigh the benefits of emission reductions due to pandemic restrictions (Wang et al., 2020; Jephcote et al., 2021).

Despite the fact that in this study we took into account the confounding effect of air temperature, wind speed and precipitation, the meteorological conditions of the accumulation and dispersion of pollutants might have not been fully addressed. This problem requires further in-depth analyses of the thermodynamic balance of the atmosphere and the advection of air masses during COVID-19 lockdowns. Further studies might also aim to increase the spatial resolution by considering both traffic and background stations from the same city, including more industrial regions like Silesia or incorporating additional pollutants, like sulphur dioxide (SO₂). It is also suggested that more complex contributions, e.g. from resuspension and secondary aerosol may be important factors influencing changes in air pollutants levels during this period (Ropkins & Tate, 2021).

The COVID-19 pandemic and the accompanying administrative restrictions have created a unique opportunity to verify whether limitation of citizens mobility and, as a consequence, urban traffic restrain is a good strategy to improve air quality in the city. Our results demonstrate that the reduction in car traffic had a positive effect only on the reduction of NOₓ concentration in Polish cities, while the overall air quality was strongly dependent on meteorological conditions and other emission sources. Moreover, many studies prove that a decrease in NOₓ concentration may result in a significant increase in tropospheric ozone level (Briz-Redón et al., 2021; Dong et al., 2021; Jephcote et al., 2021). Therefore future policy interventions to reduce emissions should take into account that limitation of individual pollutants can cause an increase in others, and may induce changes in the composition and reactivity of trace elements in the atmosphere (Wyche et al., 2021). As other studies report, during the lockdown period increases in O₃ content in troposphere enhanced the formation of secondary PM2.5 through the atmospheric oxidation of VOCs, what partially explains higher PM levels observed at that time in some places (Adam et al., 2021). Subsequently our study demonstrates trends in air pollution changes over the course of three lockdowns in Poland. Although citizens’ mobility restrictions contributed to the improvement of air quality to a small extent, this effect was not persistent. After the first impact caused by COVID-19 pandemic outburst, during the third lockdown in 2021 emissions in Polish cities returned to pre-pandemic levels. Therefore, greater remedial actions than only traffic and production restrictions will be required to achieve the EU reduction levels and the transition to the climate-neutral economy until 2050.

Editors’ note:
Unless otherwise stated, the sources of tables and figures are the authors’, on the basis of their own research.

References

Adam, M. G., Tran, P. T. M., & Balasubramanian, R. (2021). Air quality changes in cities during the COVID-19 lockdown: A critical review. *Atmospheric Research, 264*. https://doi.org/10.1016/j.atmosres.2021.105823

Baldasano, J. M. (2020). COVID-19 lockdown effects on air quality by NOₓ in the cities of Barcelona and Madrid (Spain). *Science of The Total Environment, 741*. https://doi.org/10.1016/J.SCITOTENV.2020.140353
Berman, J. D., & Ebisu, K. (2020). Changes in U.S. air pollution during the COVID-19 pandemic. *Science of the Total Environment*, 739. https://doi.org/10.1016/j.scitotenv.2020.139864

Bolaño-Ortiz, T. R., Pascual-Flores, R. M., Puliafito, S. E., Camargo-Caicedo, Y., Berná-Peña, L. L., Ruggeri, M. F., Lopez-Noreña, A. I., Tames, M. F., & Cereceda-Balic, F. (2020). Spread of covid-19, meteorological conditions and air quality in the city of buenos aires, argentina: Two facets observed during its pandemic lockdown. *Atmosphere*, 11(10). https://doi.org/10.3390/atmos11010145

Brimblecombe, P., & Lai, Y. (2020). Effect of sub-urban scale lockdown on air pollution in Beijing. *Urban Climate*, 34. https://doi.org/10.1016/j.uclim.2020.100725

Briz-Redón, Á., Belenguer-Sapiña, C., & Serrano-Aroca, Á. (2021). Changes in air pollution during COVID-19 lockdown in Spain: A multi-city study. *Journal of Environmental Sciences (China)*, 101, 16-26. https://doi.org/10.1016/j.jes.2020.07.029

Cameletti, M. (2020). The effect of corona virus lockdown on air pollution: Evidence from the city of Brescia in Lombardia region (Italy). *Atmospheric Environment*, 239. https://doi.org/10.1016/j.atmosenv.2020.117794

CDC. (2021). CDC Museum COVID-19 Timeline. https://www.cdc.gov/museum/timeline/covid19.html

Chen, Z., Hao, X., Zhang, X., & Chen, F. (2021). Have traffic restrictions improved air quality? A shock from COVID-19. *Journal of Cleaner Production*, 279. https://doi.org/10.1016/J.JCLEPRO.2020.123622

Chief Sanitary Inspectorate. (2021). Stan sanitarny kraju w 2020 roku. https://www.gov.pl/web/gis

Cichowicz, R., Wielgosiński, G., & Fetter, W. (2017). Dispersion of atmospheric air pollution in summer and winter season. *Environmental Monitoring and Assessment*, 189(12), 605. https://doi.org/10.1007/s10661-017-6319-2

Cole, M. A., Robert, ·, Elliott, J. R., & Liu, · Bowen. (2020). The impact of the Wuhan Covid-19 lockdown on air pollution and health: A machine learning and augmented synthetic control approach. *Environmental and Resource Economics*, 76, 553-580. https://doi.org/10.1007/s10640-020-00483-4

Collivignarelli, M. C., Abbà, A., Bertanza, G., Pedrazzani, R., Ricciardi, P., & Carnevale Mino, M. (2020). Lockdown for CoVId-2019 in Milan: What are the effects on air quality? *Science of the Total Environment*, 732. https://doi.org/10.1016/j.scitotenv.2020.139280

Connerton, P., Vicente de Assunção, J., Maura de Miranda, R., Dorothée Slovic, A., José Pérez-Martínez, P., & Ribeiro, H. (2020). Air quality during COVID-19 in four megacities: Lessons and challenges for public health. In *International Journal of Environmental Research and Public Health* (Vol. 17, Issue 14). https://doi.org/10.3390/ijerph17145067

Dabbour, L., Abdelhafez, E., & Hamdan, M. (2021). Effect of climatology parameters on air pollution during COVID-19 pandemic in Jordan. *Environmental Research*, 202. https://doi.org/https://doi.org/10.1016/j.envres.2021.111742

Dantas, G., Siciliano, B., França, B. B., da Silva, C. M., & Arbilla, G. (2020). The impact of COVID-19 partial lockdown on the air quality of the city of Rio de Janeiro, Brazil. *Science of The Total Environment*, 729. https://doi.org/10.1016/j.scitotenv.2020.139085

Dong, L., Chen, B., Huang, Y., Song, Z., & Yang, T. (2021). Analysis on the characteristics of air pollution in China during the COVID-19 outbreak. *Atmosphere*, 12, 205. https://doi.org/10.3390/atmos12020205

Duthel, F., Baker, J. S., & Navel, V. (2020). COVID-19 as a factor influencing air pollution? *Environmental Pollution*, 263. https://doi.org/10.1016/j.envpol.2020.114466

ECMWF. (2021). ERA5: How to calculate wind speed and wind direction from u and v components of the wind? https://confluence.ecmwf.int/pages/viewpage.action?pageId=133262398
El-Sayed, M. M. H., Elshorbany, Y. F., & Koehler, K. (2021). On the impact of the COVID-19 pandemic on air quality in Florida. *Environmental Pollution*, 285. https://doi.org/10.1016/j.envpol.2021.117451

Filonchyk, M., Hurynovich, V., & Yan, H. (2021a). Impact of (COVID)-19 Pandemic on Air Pollution in Poland Based on Surface Measurements and Satellite Data. *Aerosol and Air Quality Research*, 21(7). https://doi.org/10.4209/aaqr.200472

Filonchyk, M., Hurynovich, V., & Yan, H. (2021b). Impact of Covid-19 lockdown on air quality in the Poland, Eastern Europe. *Environmental Research*, 198. https://doi.org/10.1016/J.ENVRES.2020.110454

Fu, F., Purvis-Roberts, K. L., & Williams, B. (2020). Impact of the covid-19 pandemic lockdown on air pollution in 20 major cities around the world. *Atmosphere*, 11(11). https://doi.org/10.3390/atmos11111189

Gao, C., Li, S., Liu, M., Zhang, F., Achal, V., Tu, Y., Zhang, S., & Cai, C. (2021). Impact of the COVID-19 pandemic on air pollution in Chinese megacities from the perspective of traffic volume and meteorological factors. *Science of the Total Environment*, 773. https://doi.org/10.1016/j.scitotenv.2021.145545

Gautam, S. (2020). COVID-19: air pollution remains low as people stay at home. *Air Quality, Atmosphere and Health*, 13(7), 853-857. https://doi.org/10.1007/s11869-020-00842-6

Ginzburg, A. S., Semenov, V. A., Semutnikova, E. G., Aleshina, M. A., Zakharova, P. V., & Lezina, E. A. (2020). Impact of COVID-19 Lockdown on Air Quality in Moscow. *Doklady Earth Sciences*, 495(1), 862-866. https://doi.org/10.1134/S1028334X20110069

Gkatzelis, G. I., Gilman, J. B., Brown, S. S., Eskes, H., Gomes, A. R., Lange, A. C., McDonald, B. C., Peischl, J., Petzold, A., Thompson, C. R., & Kiendler-Scharr, A. (2021). The global impacts of COVID-19 lockdowns on urban air pollution: A critical review and recommendations. *Elementa: Science of the Anthropocene*, 9(1). https://doi.org/10.1525/elementa.2021.00176

Grzybowski, P. T., Markowicz, K. M., & Musiał, J. P. (2021a). Reduction of air pollution in Poland in spring 2020 during the lockdown caused by the covid-19 pandemic. *Remote Sensing*, 13(18). https://doi.org/10.3390/rs13183784

Grzybowski, P. T., Markowicz, K. M., & Musiał, J. P. (2021b). Reduction of air pollution in Poland in spring 2020 during the lockdown caused by the covid-19 pandemic. *Remote Sensing*, 13(18). https://doi.org/10.3390/rs13183784

Hammer, M. S., Donkelaar, A. Van, Martin, R. V., McDuffie, E. E., Lyapustin, A., Sayer, A. M., Hsu, N. C., Levy, R. C., Garay, M. J., Kalashnikova, O. V., & Kahn, R. A. (2021). Effects of COVID-19 lockdowns on fine particulate matter concentrations. *Science Advances*, 7(26). https://doi.org/10.1126/SCIADV.ABG7670

Higham, J. E., Ramírez, C. A., Green, M. A., & Morse, A. P. (2021). UK COVID-19 lockdown: 100 days of air pollution reduction? *Air Quality, Atmosphere and Health*, 14(3), 325-332. https://doi.org/10.1007/s11869-020-00937-0

Jędruszkiewicz, J., Czereńcki, B., & Marosz, M. (2017). The variability of PM10 and PM2.5 concentrations in selected Polish agglomerations: the role of meteorological conditions, 2006-2016. *International Journal of Environmental Health Research*, 27(6), 441-462. https://doi.org/10.1080/09603123.2017.1379055

Jephcote, C., Hansell, A. L., Adams, K., & Gulliver, J. (2021). Changes in air quality during COVID-19 ‘lockdown’ in the United Kingdom. *Environmental Pollution*, 272, 116011. https://doi.org/10.1016/J.ENVPOL.2020.116011

Jia, C., Fu, X., Bartelli, D., & Smith, L. (2020). Insignificant impact of the "stay-at-home" order on ambient air quality in the Memphis metropolitan area, U.S.A. *Atmosphere*, 11(6), 630. https://doi.org/10.3390/atmos11060630

Kalbarczyk, R., & Kalbarczyk, E. (2020). Meteorological conditions of the winter-time distribution of nitrogen oxides in Poznań: A proposal for a catalog of the pollutants variation. *Urban Climate*, 33. https://doi.org/10.1016/j.uclim.2020.100649
Kamińska, J. A. (2019). A random forest partition model for predicting NO₂ concentrations from traffic flow and meteorological conditions. *Science of The Total Environment*, 651, 475-483. https://doi.org/10.1016/j.scitotenv.2018.09.196

Keikhoosravi, G., & Fadavi, S. F. (2021). Impact of the inversion and air pollution on the number of patients with Covid-19 in the metropolitan city of Tehran, 37. https://doi.org/10.1016/j.juclim.2021.100867

Kerimray, A., Baimatova, N., Ibragimova, O. P., Bukenov, B., Kenessov, B., Plotitsyn, P., & Karaca, F. (2020). Assessing air quality changes in large cities during COVID-19 lockdowns: The impacts of traffic-free urban conditions in Almaty, Kazakhstan. *Science of the Total Environment*, 730. https://doi.org/10.1016/j.scitotenv.2020.139179

Kumari, P., & Toshniwal, D. (2020). Impact of lockdown on air quality over major cities across the globe during COVID-19 pandemic. *Urban Climate*, 34. https://doi.org/10.1016/j.uclim.2020.100719

Lenth, R. V. (2021). *emmeans: Estimated Marginal Means, aka Least-Squares Means*. R package version 1.7.1-1. https://cran.r-project.org/package=emmeans

Liu, F., Wang, M., & Zheng, M. (2021). Effects of COVID-19 lockdown on global air quality and health. *Science of The Total Environment*, 755. https://doi.org/10.1016/j.scitotenv.2020.142533

Marinello, S., Butturi, M. A., & Gamberini, R. (2021). How changes in human activities during the lockdown impacted air quality parameters: A review. *Environmental Progress and Sustainable Energy*, 40(4). https://doi.org/10.1002/ep.13672

Menut, L., Bessagnet, B., Siour, G., Mailler, S., Pennel, R., & Cholakian, A. (2020). Impact of lockdown measures to combat Covid-19 on air quality over western Europe. *Science of The Total Environment*, 741. https://doi.org/10.1016/J.SCITOTENV.2020.140426

Ministry of Health. (2020). *Pierwszy przypadek koronawirusa w Polsce*. https://www.gov.pl/web/zdrowie/pierwszy-przypadek-koronawirusa-w-polsce

Moreda-Piñeiro, J., Sánchez-Piñero, J., Fernández-Amado, M., Costa-Tomé, P., Gallego-Fernández, N., Piñeiro-Iglesias, M., López-Mahía, P., & Muniategui-Lorenzo, S. (2021). Evolution of gaseous and particulate pollutants in the air: What changed after five lockdown weeks at a southwest atlantic european region (northwest of spain) due to the sars-cov-2 pandemic? *Atmosphere*, 12(5). https://doi.org/10.3390/atmos12050562

Muñoz Sabater, J. (2019). *ERA5-Land hourly data from 1981 to present*. Copernicus Climate Change Service (C3S). Climate Data Store (CDS). https://doi.org/10.24381/cds.e2161bac

Nidzgorska-Lencewicz, J., & Czarnecka, M. (2015). Winter weather conditions vs. air quality in Tricity, Poland. *Theoretical and Applied Climatology*, 119(3), 611-627. https://doi.org/10.1007/s00704-014-1129-8

Onyeaka, H., Anumudu, C. K., Al-Sharify, Z. T., Egele-Godswill, E., & Mbaegbu, P. (2021). COVID-19 pandemic: A review of the global lockdown and its far-reaching effects. *Science Progress*, 104(2). https://doi.org/10.1177/00368504211019854

Parr, S., Wolshon, B., Renne, J., Murray-Tuite, P., & Kim, K. (2020). Traffic impacts of the COVID-19 pandemic: Statewide analysis of social separation and activity restriction. *Natural Hazards Review*, 21(3). https://doi.org/10.1061/(ASCE)NH.1527-6996.0000409

Piccoli, A., Agresti, V., Balzarini, A., Bedogni, M., Bonanno, R., Collino, E., Colzi, F., Lacavalla, M., Lanzani, G., Pirvano, G., Riva, F., Riva, G. M., & Toppetti, A. M. (2020). Modeling the effect of COVID-19 lockdown on mobility and NO₂ concentration in the Lombardy region. *Atmosphere*, 11(12), 1-18. https://doi.org/10.3390/atmos11121319

Pinheiro, J., & Bates, D. (2000). *Mixed-Effects Models in S and S-PLUS*. Statistics and Computing. Springer.

Pinheiro, J., Bates, D., DebRoy, S., Sarkar, D., & R Core Team. (2021). *nlme: Linear and Nonlinear Mixed Effects Models*. R package version 3.1-152. https://cran.r-project.org/package=nlme

Polednik, B. (2021a). Air quality changes in a Central European city during COVID-19 lockdown. *Sustainable Cities and Society*, 73. https://doi.org/10.1016/j.scs.2021.103096
Polednik, B. (2021b). Air quality changes in a Central European city during COVID-19 lockdown. *Sustainable Cities and Society, 73*. https://doi.org/10.1016/j.scs.2021.103096

R Core Team. (2021). *R: A Language and Environment for Statistical Computing*. R Foundation for Statistical Computing. https://www.r-project.org/

Reizer, M., & Juda-Rezler, K. (2016). Explaining the high PM10 concentrations observed in Polish urban areas. *Air Quality, Atmosphere & Health, 9*(5), 517-531. https://doi.org/10.1007/s11869-015-0358-z

Rodríguez-Urrego, D., & Rodríguez-Urrego, L. (2020). Air quality during the COVID-19: PM2.5 analysis in the 50 most polluted capital cities in the world. *Environmental Pollution, 266*. https://doi.org/10.1016/j.envpol.2020.115042

Rogulski, M., & Badyda, A. (2021). Air Pollution Observations in Selected Locations in Poland during the Lockdown Related to COVID-19. *Atmosphere, 12*(7), 806. https://doi.org/10.3390/atmos12070806

Ropkins, K., & Tate, J. E. (2021). Early observations on the impact of the COVID-19 lockdown on air quality trends across the UK. *Science of the Total Environment, 754*. https://doi.org/10.1016/j.scitotenv.2020.142374

Rossi, R., Ceccato, R., & Gastaldi, M. (2020). Effect of road traffic on air pollution. Experimental evidence from covid-19 lockdown. *Sustainability (Switzerland), 12*(21), 1-17. https://doi.org/10.3390/su12218984

Sharma, S., Zhang, M., Anshika, Gao, J., Zhang, H., & Kota, S. H. (2020). Effect of restricted emissions during COVID-19 on air quality in India. *Science of the Total Environment, 728*. https://doi.org/10.1016/j.scitotenv.2020.138878

Sicard, P., De Marco, A., Agathokleous, E., Feng, Z., Xu, X., Paoletti, E., Rodríguez, J. J. D., & Calatayud, V. (2020). Amplified ozone pollution in cities during the COVID-19 lockdown. *Science of the Total Environment, 735*. https://doi.org/10.1016/j.scitotenv.2020.139542

Silver, B., He, X., Arnold, S. R., & Spracklen, D. V. (2020). The impact of COVID-19 control measures on air quality in China. *Environmental Research Letters, 15*(8). https://doi.org/10.1088/1748-9326/ABA3A2

Singh, R. P., & Chauhan, A. (2020). Impact of lockdown on air quality in India during COVID-19 pandemic. *Air Quality, Atmosphere & Health, 13*(8), 921-928. https://doi.org/10.1007/s11869-020-00863-1

Skirienė, A. F., & Stasiškienė, Ž. (2021). COVID-19 and air pollution: Measuring pandemic impact to air quality in five European countries. *Atmosphere, 12*(3). https://doi.org/10.3390/atmos12030290

Solberg, S., Walker, S.-E., Schneider, P., & Guerreiro, C. (2021). Quantifying the impact of the Covid-19 lockdown measures on nitrogen dioxide levels throughout Europe. In *Atmosphere* (Vol. 12, Issue 2). https://doi.org/10.3390/atmos12020131

Statistics Poland. (2020). *Poland’s government agency*. https://bdl.stat.gov.pl/BDL/start

Tadano, Y. S., Potgieter-Vermaak, S., Kachba, Y. R., Chirolí, D. M. G., Casacio, L., Santos-Silva, J. C., Moreira, C. A. B., Machado, V., Alves, T. A., Siqueira, H., & Godoi, R. H. M. (2021). Dynamic model to predict the association between air quality, COVID-19 cases, and level of lockdown. *Environmental Pollution, 268*. https://doi.org/10.1016/j.envpol.2020.115920

USDT. (2010). *Public Transportation’s Role in Responding to Climate Change*. https://www.transit.dot.gov/sites/fta.dot.gov/files/docs/PublicTransportationsRoleInRespondingToClimateChange2010.pdf

Varga-Balogh, A., Leelössy, Á., & Mészáros, R. (2021). Effects of covid-induced mobility restrictions and weather conditions on air quality in Hungary. *Atmosphere, 12*(5), 561. https://doi.org/10.3390/atmos12050561

Venter, Z. S., Aunan, K., Chowdhury, S., & Leelieveld, J. (2020). COVID-19 lockdowns cause global air pollution declines. *Proceedings of the National Academy of Sciences of the United States of America, 117*(32), 18984-18990. https://doi.org/10.1073/PNAS.2006853117

Wang, P., Chen, K., Zhu, S., Wang, P., & Zhang, H. (2020). Severe air pollution events not avoided by reduced anthropogenic activities during COVID-19 outbreak. *Resources, Conservation and Recycling, 158*. https://doi.org/10.1016/j.resconrec.2020.104814
How Covid-19 pandemic influenced air quality in Polish cities – lessons from three lockdowns

Warsaw Municipal Roads Management. (2021). Little Traffic on the Roads (in Polish). https://zdm.waw.pl/aktualnosci/maly-ruch-na-drogach/

Website of the Republic of Poland. (2021). Coronavirus: information and recommendations. https://www.gov.pl/web/coronavirus/

Werner, P. A., Skrynyk, O., Porczek, M., Szczepeankowska-Bednarek, U., Olszewski, R., & Kęsik-Brodacka, M. (2021). The effects of climate and Bioclimate on COVID-19 Cases in Poland. Remote Sensing, 13(23). https://doi.org/10.3390/rs13234946

Wyche, K. P., Nichols, M., Parfitt, H., Beckett, P., Gregg, D. J., Smallbone, K. L., & Monks, P. S. (2021). Changes in ambient air quality and atmospheric composition and reactivity in the South East of the UK as a result of the COVID-19 lockdown. Science of the Total Environment, 755. https://doi.org/10.1016/j.scitotenv.2020.142526

Xu, Y., Zhu, B., Shi, S., & Huang, Y. (2019). Two inversion layers and their impacts on PM2.5 concentration over the Yangtze River Delta, China. Journal of Applied Meteorology and Climatology, 58(11), 2349-2362. https://doi.org/10.1175/JAMC-D-19-0008.1

You, Y., Byrne, B., Colebatch, O., Mittermeier, R. L., Vogel, F., & Strong, K. (2021). Quantifying the impact of the covid-19 pandemic restrictions on CO, CO₂, and CH₄ in downtown toronto using open-path fourier transform spectroscopy. Atmosphere, 12(7). https://doi.org/10.3390/atmos12070848

Zangari, S., Hill, D. T., Charette, A. T., & Mirowsky, J. E. (2020). Air quality changes in New York City during the COVID-19 pandemic. Science of The Total Environment, 742, 140496. https://doi.org/10.1016/J.SCITOTENV.2020.140496

Zhu, Z., Qiao, Y., Liu, · Quyue, Lin, C., Dang, E., Fu, W., Wang, G., & Dong, J. (2021). The impact of meteorological conditions on Air Quality Index under different urbanization gradients: A case from Taipei. Environment, Development and Sustainability, 23, 3994-4010. https://doi.org/10.1007/s10668-020-00753-7
