Ranking Three Water Sensitive Urban Design (WSUD) Practices Based on Hydraulic and Water Quality Treatment Performance: Implications for Effective Stormwater Treatment Design

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Abstract: Bioretention basins, constructed wetlands and roadside swales are among the most common Water-Sensitive Urban Design (WSUD) or stormwater quality treatment systems. Although these systems can reduce stormwater quantity and improve quality, their hydraulic and water quality treatment performances are different. The aim of this study was to investigate the hydraulic and water quality performance of a bioretention basin, a constructed wetland and a roadside swale by analyzing monitored water quantity and quality data from a range of rainfall events using a ranking approach. The study outcomes showed that a bioretention basin performed better in relation to peak flow and runoff volume reduction while the constructed wetland tended to produce better outflow water quality. The roadside swale had a relatively lower capacity for treating stormwater. These results suggest that a bioretention basin could be the preferred option when the primary requirement is water quantity improvement. However, if water quality improvement is the primary consideration, a constructed wetland could be more efficient. Additionally, when designing a treatment train, it appears to be preferable to place a bioretention basin prior to a constructed wetland. Further, a swale appears to be more appropriate for use as a pretreatment device. The research study outcomes will contribute to effective stormwater treatment design.

Keywords: bioretention basin; constructed wetland; roadside swale; Water-Sensitive Urban Design (WSUD); stormwater quality; stormwater pollutant processes

1. Introduction

Urbanisation results in an increase in impervious surfaces and the occurrence of diverse anthropogenic activities common to such areas [1–3]. These lead to increased stormwater runoff volume, peak flow and flow velocities and the introduction of a range of physical, chemical and biological pollutants, resulting in the degradation of stormwater quality and increased pollutant loads to receiving water bodies [4–10]. Water-Sensitive Urban Design (WSUD) measures are commonly implemented to mitigate quantity and quality impacts in relation to stormwater runoff from urban areas [11–13]. WSUDs primarily use vegetation and retention/detention systems placed close to stormwater source areas to reduce peak flow and runoff volume and to remove pollutants [14–16].

Bioretention basins, constructed wetlands and swales are among the most common WSUD measures used worldwide [16–20]. Bioretention basins operate by passing stormwater runoff through prescribed filter media with planted vegetation [18,21]. Therefore, bioretention basins treat stormwater by both, vegetation and filter media. Stormwater runoff reaching a bioretention basin will infiltrate and percolate through the filter media, while excessive flow can bypass the system. Generally, the stormwater runoff does not have high velocity when flowing through a bioretention basin due to the design features. Constructed wetlands are artificial, shallow and extensively vegetated water bodies. They
aim to remove pollutants including dissolved pollutants from stormwater runoff, enhance landscape amenity as well as ensure stormwater reuse as a supplementary benefit [22,23]. Commonly, a constructed wetland contains an inlet zone, a macrophyte zone and a bypass channel. Roadside swales are vegetated strips with a shallow channel and commonly located along a road. Roadside swales disconnect impervious surfaces from downstream waterways and are generally designed with side slopes. Roadside swales can reduce flow velocities and remove large particles from stormwater runoff [24,25].

Although all three treatment measures are capable of mitigating stormwater quantity impacts, such as runoff peak and volume, and improving quality, their hydraulic and water quality performances can be different due to the involvement of different mechanisms and the influence exerted by rainfall–runoff characteristics [26,27]. For example, the treatment efficiency of a bioretention basin is significantly influenced by the antecedent dry period [26], while rainfall depth is the primary influential factor in the case of constructed wetlands [27]. Additionally, the placements of WSUDs are generally different. For instance, swales are commonly placed along roadsides, while bioretention basins and constructed wetlands require relatively larger land extents. In this regard, selecting appropriate WSUDs for a given urban area to ensure effective stormwater treatment is a challenging task. Furthermore, since each WSUD has varied performance characteristics in relation to the treatment of stormwater, a treatment train is commonly adopted to increase stormwater treatment efficiency. A treatment train refers to a combination of WSUD measures in series for effective flow volume reduction and pollutant removal [28].

For treatment system design, it is essential to define the pollutant loads and/or concentrations to be reduced prior to stormwater entering receiving waters in order to achieve specific water quality objectives [29,30]. In this context, in-depth knowledge of the treatment efficiency of commonly used WSUDs is essential for appropriate system selection and their placement in order to guide the treatment strategy.

Accordingly, the aim of this study was to investigate the hydraulic and water quality performance of a bioretention basin, a constructed wetland and a roadside swale by analyzing monitored water quantity and quality data for a range of rainfall events. The scientific question addressed is: What is the ranking of these systems in relation to their hydraulic and water quality treatment efficiency? The research outcomes will contribute to effective stormwater treatment design.

2. Materials and Methods
2.1. Study Sites

Coomera Waters was selected as the study site. This is a newly developed residential estate in Gold Coast City, Australia. The residential area includes a bioretention basin (B), a constructed wetland (W) and roadside swales (S). The study site location and the three stormwater treatment systems in the field are shown in Figure 1. Detailed information about the three WSUD systems, including engineered characteristics, is provided in the Supplementary Materials.
Figure 1. Study site. TA: total area; IA: impervious surface area; IA%: impervious surface percentage. Stormwater runoff from sub-area 1 primarily contributes to B. After passing through B, the stormwater enters the constructed wetland. Stormwater runoff from sub-area 2 and sub-area 3 directly flows to W. Sub-area 4 distributes flow to roadside swales, while sub-area 5 is the bypass of W, which does not receive treatment from any of the WSUDs.

2.2. Data Collection

2.2.1. Monitoring Program

The inlets and outlets of the three systems (B, W and S) were monitored using automatic monitoring stations. These monitoring stations recorded rainfall–runoff data and captured time-based stormwater runoff samples for water quality parameter testing. Flow measurements were conducted using calibrated V-notch weirs, while stormwater samples were collected using stage triggered, peristaltic pumping. Accordingly, flow measurements from 22 events for B, 17 events for W and 13 events for S were collected, while the corresponding values for water quality data were 12 events for B, 7 events for W and 7 events for S. The primary reason for the inconsistency in rainfall event numbers was due to equipment malfunction. The characteristics of the rainfall events monitored were, 1.0–93.6 mm depth, which are within the 1-year ARI (Average Recurrence Interval) range. The 1-year ARI is the rainfall range on which most WSUD system designs are based [31]. For each event, 10–24 samples were collected using an automatic sampler, according to the rainfall characteristics. The characteristics of the rainfall events monitored, including sampling dates, are provided in Table S1 in the Supplementary Materials.

2.2.2. Parameter Selection and Laboratory Testing

The peak flow and runoff volume at inlets and outlets were recorded as the primary hydraulic characteristics [21,32]. Based on this, the peak and volume reduction percentages were determined and considered as important parameters representing the quantity mitigation performance of the WSUDs. In terms of water quality, the samples collected were tested for total suspended solids (TSS), total phosphorus (TP) and total nitrogen (TN).
The testing methods used were according to the Standard Methods for the Examination of Water and Wastewater [33]. The detailed information about TSS, TN and TP testing is provided in the Supplementary Materials. TN, TP and TSS were selected due to the fact that they are among the primary stormwater pollutants [6,34]. Additionally, total pollutant loads as well as event mean concentrations (EMCs) were determined for each selected rainfall event.

2.2.3. An Assumption Made When Using Outflow Water Quality

In terms of S, the inflowing stormwater flows along both sides of the road (see Figure 1). This resulted in difficulty in collecting inflow stormwater samples from S. Due to this limitation, it was not possible to compare water quality between the inlet and outlet of S. Therefore, a water quality analysis was undertaken by comparing outflow pollutant EMCs and loads from the three WSUDs. EMC data and loads in relation to the three WSUDs are provided in Table S2 in the Supplementary Materials. The analysis was based on the premise that the inflow stormwater quality into B and W is similar to the inflow stormwater quality into S. This was considered reasonable since these three systems are located close to each other and all receive stormwater from the Coomera Waters catchment with similar land use characteristics (see Figure 1).

2.3. Data Analyses

The study approach included two primary steps. Step 1 was to initially compare the hydraulic and water quality treatment performances of the three systems (B, W and S). This was conducted using boxplots. Step 2 was to rank the three WSUDs. The rankings were initially undertaken for hydraulic performance, pollutant EMCs and pollutant loads, separately. Thus, it was possible to gain an individual comparison of the different treatment systems. For example, only hydraulic performance needs to be understood if a WSUD is only designed for quantity mitigation. However, if a WSUD is also expected to undertake quantity and quality mitigation, knowledge of hydraulic and water quality treatment performance is required. Secondly, comprehensive ranking was conducted by taking into account both, water quantity and quality treatment performance.

The PROMETHEE method was used for the data analysis because of its capability to rank objects with a range of variables [35]. In PROMETHEE, a ranking order is developed according to the net ranking flow, the $\Phi$ values, for a number of objects (the WSUDs) on the basis of a range of variables (hydraulic and water quality performance). The $\Phi$ values are calculated for each object based on the partial ranking outflow indices, $+\Phi$ and $-\Phi$. The objects are rank-ordered from the most preferred (the most positive ($+$) $\Phi$ value) to the one with the lowest performance (the most negative ($-$) $\Phi$ value). A large difference between two net ranking outflow values, $\Phi$ values, indicates that the two objects are dissimilar. Detailed information regarding the PROMETHEE method is provided in the Supplementary Materials. It can also be found in Keller et al. [36] and Khalil et al. [37].

3. Results and Discussions

3.1. Analysis of Hydraulic and Water Quality Treatment Performance

3.1.1. Hydraulic Performance

Figure 2 compares the peak flow and runoff volume reduction percentages of B, W and S. It was noted that B and W showed higher peak flow reduction percentages than S. The average peak flow reduction percentages were 94.2% and 98.7% for B and W, respectively, while the corresponding value was 63.2% for S. Additionally, B and W indicated more consistent peak flow reduction performance than S since the data ranges of B and W are much narrower than for S (see Figure 2a). In terms of runoff volume reduction percentage (see Figure 2b), B shows a slightly higher value than the other two systems, since the average runoff volume reduction percentages were 49.5%, 29.6% and 34.1% for B, W and S, respectively. However, all three systems have wide ranges of runoff volume reduction
percentages. This means that the volume reduction efficiency of the three systems varies highly with rainfall characteristics as the data were collected for a range of rainfall events.

Figure 2. Hydraulic performance of WSUDs, (a) peak flow reduction; (b) flow volume reduction. B = bioretention basin; W = constructed wetland; S = roadside swale.

3.1.2. Water Quality Performance

Figure 3 shows the outflow pollutant EMCs of B, W and S (Figure 3a for TSS, Figure 3b for TN and Figure 3c for TP). It is evident that B has the highest outflow of pollutant EMCs (the average values were 28.4 mg/L for TSS, 1.49 mg/L for TN and 0.14 mg/L for TP) regardless of the pollutant species, while W has the lowest outflow EMC values for TN and TP. W and S have a relatively similar outflow of TSS EMCs. These results imply that the bioretention basin is relatively less effective in terms of pollutant EMC reduction than the wetland and swale, since the bioretention basin produced the highest pollutant EMCs in the outflow of. It is also noteworthy that the outflow pollutant EMCs for B have a wider data range than for the other two systems. This indicates the high variability of outflow water quality from the bioretention basin compared to the constructed wetland and roadside swale.

Figure 3. Outflow pollutant EMCs of WSUDs, (a) TSS; (b) TN; (c) TP. B = bioretention basin; W = constructed wetland; S = roadside swale.

In the case of outflow pollutant loads (see Figure 4), S shows the highest average values for TSS (0.57 kg/ha), TN (0.03 kg/ha) and TP (0.003 kg/ha) among the three WSUD
systems, while W is the lowest for all three pollutants. Additionally, the data range of S is the widest, followed by B, while W has the narrowest data range. This means that the outflow pollutant loads from the constructed wetland are relatively consistent, having lower variability, compared to the other two systems. These observations imply that the constructed wetland performs better in terms of removing pollutant loads since it has the lowest and the most consistent outflow pollutant loads, while the roadside swale is the least effective among the three WSUDs as it has higher outflow pollutant loads as well as the highest variation.

Figure 4. Outflow pollutant loads of WSUDs, (a) TSS; (b) TN; (c) TP. B = bioretention basin; W = constructed wetland; S = roadside swale.

3.2. Ranking of WSUD Systems

As discussed above, the three WSUD systems showed significant differences in treatment performance in terms of water quantity and quality. For example, the bioretention basin exhibited good hydraulic performance while the constructed wetland demonstrated relatively better water quality treatment performance. These differences in performance among the WSUDs make it difficult to directly compare their overall stormwater treatment capabilities, particularly when considering both quantity and quality mitigation. This results in the inability to provide technically robust guidance in relation to the selection of appropriate WSUDs and their arrangement in a treatment train. In this context, a comprehensive approach to ranking these systems based on their hydraulic and water quality treatment efficiencies needs to be employed.

The parameters used for ranking were average values and coefficient of variation (CV) values for peak flow reduction, runoff volume reduction, TSS, TN and TP outflow EMCs and loads for rainfall events monitored for each WSUD. For example, there were 22 rainfall events monitored for peak flow reduction in the bioretention basin (see Section 2.2.). Therefore, parameters used for ranking were the average and CV values of peak flow reduction percentages for the 22 rainfall events. The same method was applied for the other WSUDs to assess their hydraulic and water quality treatment performance. The CV represents the variability in data and was obtained from the ratio of the standard deviation and the average values. The reason for considering both average and CV values
was because not only treatment efficiency but also stability in treatment performance (representing the variability in treatment performance for different rainfall events) plays an important role in representing a WSUD’s performance. Accordingly, three matrices were generated, including $3 \times 4$ (hydraulic performance: average peak flow reduction, average runoff volume reduction, peak flow reduction CV value and runoff volume reduction CV value for the three WSUDs), $3 \times 6$ (outflow pollutant EMCs: average TSS, TN and TP outflow EMCs and their CV values for the three WSUDs) and $3 \times 6$ (outflow pollutant loads: average TSS, TN and TP outflow loads and their CV values for the three WSUDs).

For PROMETHEE ranking, three matrices were separately developed, while comprehensive ranking was undertaken using data from all three matrices together. The PROMETHEE analysis requires: (i) an assignment of the ranking sense (maximise or minimise); (ii) the selection of a preference function from Linear, V-shape and Usual functions; and (iii) the weighting of each variable based on importance. In this case, average peak flow reduction and runoff volume reduction were maximised, while all CV values and outflow pollutant EMCs and loads were minimised. This was to top-rank the WSUD system with high peak and volume reductions, low outflow pollutant amounts and low outflow quantity and quality variability. The V-shape preference function was chosen because it is commonly applied in stormwater quality research [38,39]. Furthermore, each variable was equally weighted since they were considered to be equally important. Accordingly, the resulting ranking is given in Table 1.

**Table 1.** Ranking results. B = bioretention basin; W = constructed wetland; S = roadside swale.

| Treatment Performance | WSUD System | $\Phi$ Value | Ranking |
|-----------------------|-------------|--------------|---------|
| Hydraulic performance | B           | 0.2500       | 1       |
|                       | W           | 0.1250       | 2       |
|                       | S           | −0.3750      | 3       |
| Outflow pollutant EMCs| W           | 0.1667       | 1       |
|                       | S           | −0.0833      | 2       |
|                       | B           | −0.0834      | 3       |
| Outflow pollutant loads| W          | 0.4167       | 1       |
|                        | B           | 0.0833       | 2       |
|                        | S           | −0.5000      | 3       |
| Hydraulics and water quality combined | W | 0.2188 | 1 |
|                                | B           | 0.0938       | 2       |
|                                | S           | −0.3125      | 3       |

As shown in Table 1, the three WSUDs have different rankings based on hydraulic and water quality treatment performance. For hydraulic performance, the ranking was $B > W > S$, which indicates that the bioretention basin performed best in terms of peak and volume reduction. For outflow EMCs, the ranking was $W > S \approx B$ (S and B have similar $\Phi$ values), while it was $W > B > S$ for outflow pollutant loads. This means that the constructed wetland shows the best pollutant treatment efficiency compared to the bioretention basin and roadside swale. Considering both hydraulic and water quality treatment performance, the ranking was $W > B > S$. This implies that the constructed wetland has the best treatment performance in relation to both water quantity and quality. These results suggest that the bioretention basin should be the preferred option when the primary design focus is water quantity. However, if the focus is water quality improvement, the constructed wetland would be more efficient. Furthermore, in the case of a treatment train, a bioretention basin should be placed prior to a constructed wetland. This is because the bioretention basin would initially reduce the runoff volume as it has better hydraulic performance, resulting in reduced inflow into the constructed wetland. This is also supported by the fact that a smaller runoff volume can generally enhance treatment performances in a constructed wetland [31]. In the case of the roadside swale, due to the relatively lower capacity for treating stormwater, it can be used as a pretreatment system.
prior to the stormwater flow into other WSUD devices, such as a bioretention basin and/or a constructed wetland.

4. Conclusions

This research study investigated three typical WSUDs, namely, a bioretention basin, a constructed wetland and a roadside swale. Both, water quantity and quality treatment capacities were compared using a ranking analysis. The ranking took into consideration treatment efficiency and treatment stability (representing the variability in treatment performance with different rainfall events). The three WSUDs exhibited different characteristics in the mitigation of water quantity and quality. The bioretention basin showed better performance in relation to peak flow and runoff volume reduction, while the constructed wetland tended to produce better outflow water quality. These results may imply that, when designing a treatment train, it is important to place a bioretention basin prior to a constructed wetland. This is due to the better hydraulic performance of a bioretention basin which could initially reduce runoff volume, resulting in smaller inflow into the constructed wetland. The stormwater treatment performance of the roadside swale was comparatively low. Hence, it might be more appropriate for use as a pretreatment device. These research outcomes can provide important guidance for stormwater treatment strategy design and are expected to contribute to effective stormwater management. However, it should be noted that additional investigations are needed to further confirm these results by monitoring more rainfall events with complete hydraulic and water quality data. Additionally, it is also recommended for future research that more WSUD systems should be investigated in order to achieve a more in-depth understanding of their water quantity and quality treatment performance. These additional studies will provide robust guidance to system designers for the selection of appropriate treatment systems and their placement for effective performance.

Supplementary Materials: The following supporting information can be downloaded at https://www.mdpi.com/article/10.3390/w14081296/s1, Table S1: Characteristics of rainfall events selected and hydraulic performance of the bioretention basin, constructed wetland and roadside swale; Table S2: Water quality performance of the bioretention basin, constructed wetland and roadside swale; Pollutant testing methods; PROMETHEE method; Information about the three WSUD systems. References [40,41] are cited in the supplementary materials.

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