Reduced Water Negatively Impacts Social Bee Survival and Productivity Via Shifts in Floral Nutrition

Erin E. Wilson Rankin,1,2,* Sarah K. Barney,1 and Giselle E. Lozano1

1Department of Entomology, University of California, Riverside, 900 University Ave, Riverside, CA 92521 and 2Corresponding author, e-mail: e.wilson.rankin@gmail.com

Subject Editor: Johanne Brunet

Received 17 February 2020; Editorial decision 6 September 2020

Abstract

Pollinators provide a key ecosystem service vital for the survival and stability of the biosphere. Identifying factors influencing the plant-pollinator mutualism and pollinator management is necessary for maintaining a healthy ecosystem. Since healthy beehives require substantial amounts of carbohydrates (nectar) and protein (pollen) from forage plants such as clover, we must assess how resources offered by plants change under limited water conditions in order to fully understand how drought modifies the pollination mutualism. Here we document how reduced water availability leads to decreased nectar quality and quantity and decreased protein quality of pollen. Furthermore, we provide conclusive evidence that these lower quality resources lead to decreased survival and productivity in both developing honey bees (Hymenoptera: Apidae) and bumble bees (Hymenoptera: Apidae). The results emphasize the importance of the nutritional effects of reduced water on bees when predicting shifts of pollination mutualisms under climate change.

Key words: water availability, nutrition, pollinator, foraging, drought

Bees are some of the most economically important beneficial insects in the United States (James and Pitts-Singer 2008). Because of the vital ecological economic importance of their pollination services, maximizing bee health is of paramount importance. Stressors at the ecosystem-level can affect both plant reproduction and pollinator health (Cayan et al. 2008). One such emerging threat is climate change, which includes multiple stressors, such as warmer temperatures and increasing incidence of drought (Intergovernmental Panel on Climate Change 2002, Phillips et al. 2018). Shifts in temperature or precipitation can significantly alter plant physiology, phenology (Penuelas et al. 2004, Andresen et al. 2010, Brunet and Larson-Rabin 2012) and survival (see references in McDowell et al. 2013). Because bees are obligate florivores and acquire most of their nutrients from nectar and pollen (Jordano et al. 2006, Winfree 2010), any altered availability and quality of bee forage may affect bee health (Di Pasquale et al. 2013, Ruedenauer et al. 2015, Smart et al. 2019), with cascading effects on pollination services.

Diet affects bee development and health, particularly immunocompetency (Alaux et al. 2010), and bees require a rich and diverse diet for optimal health (Di Pasquale et al. 2013). Therefore, to investigate how reduced water conditions modify the pollination mutualism, we must first quantify how reduced water affects the resources offered by the plant, and second, measure how such changes in floral nutritive value affect bee health and productivity. Under severe drought stress or extreme desiccation or dehydration strain as defined by Blum and Tuberosa (2018), plants may abort flowers entirely (Peterson et al. 1993, Fang et al. 2010) or senesce (McDowell 2011). However, plants experiencing limited water conditions do not necessarily experience desiccation; rather, plants under low, but not dehydration-inducing conditions may produce fewer (Burkle and Runyon 2016, Phillips et al. 2018) or smaller flowers (Gallagher and Campbell 2017). Although, how the nutritive value of floral resources is influenced by reduced water has received little attention.

Changes in floral resource availability and quality bring about shifts in foraging behavior (Harder 1986) and floral selection (Cnaani et al. 2006). Recent studies found that variations in water available to plants led to altered pollinator visitation rates (Gallagher and Campbell 2017, Descamps et al. 2018), as well as detectable changes in nectar volume and corolla size (Gallagher and Campbell 2017, Descamps et al. 2018, Brunet and Van Etten 2019). Such changes in direct response to water availability may lead to a morphological mismatch between corolla length and pollinator tongue length (Gérard et al. 2020). While there have been studies documenting the negative effects of drought on nectar volume and sugars (e.g., Waser and Price 2016, Descamps et al. 2018), little is known about how water limitation affects the nutritive qualities and availability of both nectar and pollen. Furthermore, no studies to date directly link such reductions in floral rewards to decreases in bee development and productivity.

Here, we use a combination of manipulative experiments to follow and quantify the direct impacts of water limitation on the quantity and quality of floral rewards and to assess the subsequent impacts on bee development and colony productivity. We test the hypothesis...
that nutrition available to honey bees and bumble bees is reduced when plants have less water available yet are not experiencing water stress. Understanding the impacts of reduced water availability on pollen and nectar production and consequences for bee health will help develop recommendations to maximize pollination services and nutrient availability under restricted water or drought conditions.

Materials and Methods

Water Limitation Impacts on Nectar and Pollen

We measured the quantity and quality of floral nectar and pollen produced by florets under optimum and lower water conditions. For the experimental trials, we focused on a common forage plant native to western North America, tomatc clover (Trifolium willdenovii). Seeds were obtained from Larner Seeds (Bolinas, CA), scarified, and germinated in germination trays using UC Soil Mix III (Supp Table S1 [online only]). For each experiment we grew 160 plants in Deepots (D40, Stuewe & Sons) in a temperature-controlled greenhouse at the University of California, Riverside: 80 plants experienced optimal water conditions and 80 plants experienced reduced water conditions (drought treatment). All plants were watered using the greenhouse’s fertilized water system (Peters Excel 25.5-5-20 at 1:100 injection rate, Scotts Miracle-Gro Company, Marysville, OH). Optimal water conditions were assessed by determining how much water was required for the soil to be at field capacity (Colman 1947). The reduced water treatment was watered just as often as the optimal treatment but received 30% less water than the optimal treatment. This reduction corresponded with the median required water was required for the soil to be at field capacity (Colman 1947). The reduced water treatment was watered just as often as the optimal treatment but received 30% less water than the optimal treatment. This reduction corresponded with the median required water for the soil to be at field capacity (Colman 1947).

To collect nectar and pollen, we removed entire inflorescences one day after the first florets opened, and we counted the number of open florets for each inflorescence. For each open floret, we used microscissors to separate the anthers from the filaments and set these aside for pollen collection (see below). Each floret was then pinned to a cork stopper (Kearns and Inouye 1993). This floret was then concentrated in %BRIX using a refractometer. For most nectar samples, we used a hand-held refractometer (Bellingham and Stanley, Eclipse 45-81); however, for the few nectar samples with very high %BRIX (>50%) we used a refractometer designed for high sugar content (Ade Advanced Optics, RHW-80ATC).

In addition to harvesting nectar from each plant during its bloom period, we collected the anthers for each floret (Supp Material [online only]). We subsequently separated the pollen from anthers using a microcentrifuge, and pooled pollen by plant in a pre-weighed microcentrifuge tube. We collected pollen from an average of 85.8 ± 22.6 florets per plant and weighed pollen using a Sartorius M2P balance. To assess pollen quality, we divided the total mass of pollen collected per plant by the number of florets collected yielding the average mg of pollen per floret for each plant. To analyze pollen quality, we extracted total protein using a protocol adapted from Vanderplanck et al. (2014) and Wang et al. (2006) (Supp Material [online only]), then quantified total protein using a bicinchoninic acid (BCA) Protein Assay following manufacturer’s directions (Pierce, Thermo Scientific).

To determine if reduced water affected plant fitness, we grew an additional 80 plants (40 plants under each of the two watering conditions), hand self-pollinated the seven to 10 florets per inflorescence per plant using a camel-hair paintbrush and quantified the resulting seed set (e.g., number of seeds per floret and seed weight). Pollen and nectar collected from these 80 plants were included in the assessments of floral reward quantity and quality described above. At the time of seed collection, we harvested, dried, and weighed each plant to obtain total plant biomass.

Water Limitation Impacts on Bee Survival Via Changes in Floral Resources

To link the impacts of reduced water availability to plants directly to bee survival and productivity, we developed artificial diets based on the nutritional composition offered by clover inflorescences. Using a sucrose/glucose/fructose assay kit (Megazyme Bray, Ireland), we found a 50:23:27 sucrose:glucose:fructose ratio in clover nectar in both optimally treated and reduced water treatments. Pilot experiments revealed that nectar from reduced water plants had 17% lower sugar concentration (BRIX) as compared to the nectar of optimum plants and that pollen from reduced water treatment had 10% less protein than pollen of optimum watered plants. Based on the differences in our nutritional analyses of clover described above, we manipulated the protein quality of bee diets by manipulating the amount of royal jelly (Crockett Honey, Temple AZ) provided to honey bees or the pollen (Brushy Mountain, Moravian Farms, NC) provided to bumble bees. Quality of artificial diets were double checked using a refractometer for nectar and the assay kits described above for pollen to ensure the appropriate differences in quality for the two treatments. These diets were provided to developing larval bees (Apis mellifera), bumble bee microcolonies (Bombus impatiens) (Cnaani et al. 2002) or to bumble bee colonies (B. impatiens).

Effect of Nutrition on Developing Honey Bees

Following the methods of Di et al. (2016), Aupinel et al. (2005), and Peng et al. (1992), we obtained 1-d-old larvae from an Apis mellifera colony at the University of California, Riverside apiary. Using grafting tools (Sinova, Zhengzhou, China), we grafted 96 one-day-old honey bee larvae onto 250 μl of artificial diet in queen cell cups. Cell cups were then placed into 48-well plates (Corning, New York, NY), covered, and maintained in a dark environmental chamber (34 ± 1°C, 95 ± 0.5% humidity). All larvae were assessed for four hours and 24 h after grafting to determine the rate of mortality due to grafting. Diets were modified from Di et al. (2016) to reflect our observed 10% and 17% reductions in protein and carbohydrates, respectively, as a result of water limitation to clover. The optimum water diet consisted of 53% (w/w) royal jelly, 6.5% sucrose, 3.1% glucose, 3.4% fructose, and 34% water, while the reduced water diet was comprised of 47.6% royal jelly, 5.4% sucrose, 2.6% glucose, 3.4% fructose, and 41.5% water. To reflect the reduced protein available to developing larvae, we altered the concentration of royal jelly in the larval provisions. Forty-eight larvae received the optimum diet and 48 larvae received a diet reflecting the relative carbohydrate and protein concentrations of the reduced water plants. Bees were observed between 0900 and 1000 hours every morning, noting growth, onset of pupation, adult emergence, and mortality. All adults were weighed at the end of the experiment.
Effect of Nutrition on Bumble Bees
In addition to honey bees, we raised Bombus impatiens on each of the two diets based off the nutritional values of the clover grown under the optimal or the reduced water conditions. We created and maintained 74 microcolonies each consisting of three workers from the same source colony, on these artificial diets (i.e., 37 microcolonies per treatment) using six source colonies. Small, queenless microcolonies are commonly used to study bumble bee development and survival (Caan et al. 2002, Moermon et al. 2017) and allows for treatments to be balanced across source colonies (Molgarejo et al. 2018). To reflect the 10% reduction in protein and 17% reduction in carbohydrates due to water reduction in clover, we created the following diets. The optimum water diet consisted of 0.250 g pollen and a sugar solution of 29.4% sucrose, 13.5% glucose, 15.9% fructose, and 41.2% water, while the reduced water diet included 0.225 g pollen and a sugar solution of 24.8% sucrose, 11.4% glucose, 13.4% fructose, and 50.6% water. All bumble bee workers were marked with unique number tags to allow for individual identification and calculation of forager lifespans and mortality rates. Fresh food was provided daily. We also noted the creation of honey pots and any incidence of worker egg laying in the microcolonies.

We also raised entire Bombus impatiens colonies (n = 4 colonies total, two colonies per treatment) ad libitum on these artificial diets. At the beginning of the experiment, each colony had one queen and 40–50 workers. Similar to microcolonies, all workers were individually marked. We recorded worker body mass, survival and mortality, as well as new worker (i.e., callow) and gyno production daily.

Statistical Analyses
All analyses were conducted in R v. 3.6.0 (R Core Team 2020) and means are reported a standard error. All generalized linear mixed-effects models (GLMMs) were conducted using the lme4 package (Bates et al. 2013). Gaussian error distribution was used unless otherwise noted. All plant phenological variables (time from scarification to first floret, number of inflorescences, number of florets per inflorescence, total number of florets, total plant dry biomass, and seed weight) were analyzed using GLMMs where treatment was the fixed effect and random effects were plant ID and sample collection date. Plant productivity (seed set per inflorescence and the number of pollinated florets per inflorescence that set seed) was similarly assessed using a Poisson error structure. To assess the effects of water treatment on the quantity or quality of nectar or pollen on a per floret basis, we used GLMMs with treatment as the fixed effect, plant ID and sample collection date as random effects. To assess treatment effects on the quantity or sugar concentration of nectar and the quantity and protein concentration of pollen on a per plant basis, we used GLMMs with treatment and number of florets as fixed effects, sample collection date as a random effect.

For each bee species separately, survival curves were estimated for both diets using the Kaplan-Meier method and compared using the ‘survdiff’ function in package survival (Therneau 2015). For honey bees, we used χ² tests to compare number of bees reaching each stage of development and mortality before adulthood between treatments. In package stats (R Core Team 2020), we used t-tests to assess treatment effects on development time and body mass of bees. For Bombus microcolonies, a GLMM examined the effect of treatment on forager body mass with microcolony identity as a random effect. We calculated odds ratio (OR) and Pearson χ² for the creation of honey pots and incidence of egg laying across microcolonies by diet. For full bumble bee colonies, we used χ² tests to compare the total numbers of new workers and gynes produced between the treatments.

Results
Water Limitation Impacts on Nectar and Pollen
Watering regimes altered the phenology of clover and the quality of its floral rewards. Reduced water led to increased time until first floret opening by 3 d (F1,71 = 4.70, P = 0.034) and resulted in half as many inflorescences per plant as compared to the optimal water treatment (Fig. 1a: F1,100 = 73.72, P < 0.00001). Plants in the optimal treatment produced significantly more florets per inflorescence (12.0 ± 0.3 florets) than those in the reduced water treatment (9.9 ± 0.3 florets) (F1,4 = 18.45, P = 0.0023). Moreover, optimal plants produced an average of 165.4 ± 15.8 total florets, while reduced water plants had an average of 118.6 ± 5.8 total florets. Watering regime also had an effect on total plant biomass: plants in the optimal watering treatment had 46% more dry plant biomass than plants in the reduced water treatment (F1,31 = 19.36, P < 0.0001).

Watering treatment also affected seed set: the number of seeds produced per plant were more than twofold higher in the optimal treatment as compared to the reduced water treatment (Fig. 1b: F1,31 = 11.64, P = 0.0092). Across treatments, the percentage of florets per inflorescence that set seed were similar (reduced water: 48.6% and optimal: 59.9%); however, the total number of seeds produced per inflorescence did differ among treatments (F1,615 = 16.05, P < 0.0001), with optimal plants producing 23% more seeds per inflorescence (5.9 ± 0.3 seeds vs 4.7 ± 0.2 seeds, respectively). The weight of an individual seed, however, was unaffected by treatment (reduced water: 2.91 ± 0.2 mg and optimal: 3.04 ± 0.2 mg; F1,31 = 0.0072, P = 0.93).

Fig. 1. Reduced water impacts plant productivity. (a) Number of inflorescences per clover plant when grown under optimal and 30% reduced water conditions. (b) Number of seeds per clover plant produced under optimal and 30% reduced water conditions.
Nectar Quantity and Quality
Inflorescences from optimal water plants produced with an average of 0.09 ± 0.01 µl of nectar per floret while inflorescences from low water plants exhibited an average of 0.03 ± 0.01 µl of nectar per floret ($F_{1,125} = 7.86, P = 0.0059$). Consequently, we observed that higher water treatment led to increased nectar volumes overall as compared to the reduced water treatment ($F_{1,125} = 12.02, P = 0.0007$). Moreover, nectar from optimal plants exhibited 11% higher sugar concentration (BRIX) than nectar from reduced water plants ($F_{1,66} = 12.38, P = 0.0008$).

Pollen Quantity and Quality
The mass of pollen per floret did not differ between treatments: 0.076 ± 0.007 mg for optimal water plants and 0.085 ± 0.01 mg for reduced water plants ($F_{1,114} = 1.27, P = 0.26$). Importantly, the total amount of pollen produced per plant decreased 36% by the water treatment ($F_{1,114} = 4.87, P = 0.029$), reflecting the difference in total number of florets by treatment. Analysis of total raw protein for these pollen samples revealed 8% more protein in the optimal treatment than pollen from the reduced water treatment ($F_{1,28} = 4.97, P = 0.034$).

Water Limitation Impacts on Bee Survival Via Changes in Floral Resources
Effect of Nutrition on Developing Honey Bees
Mortality due to the grafting process did not differ between the treatments (optimal: 19% vs reduced water: 15%). There was a significant effect of diet quality on honey bee survival post-grafting (Kaplan-Meier: $X^2 = 6.9, P = 0.009$), where bees in the optimal diet treatment had 26% higher survivorship than those on the reduced water diet. Nutrition also had a significant impact on total development time from egg until adult emergence ($t_{13} = 3.42, P = 0.0046$). Individuals raised on the optimal diet emerged as adults on 7.4 ± 0.2 d after pupating, whereas those from the reduced water diet emerged as adults 8.6 ± 0.3 d after beginning pupation ($t_{13} = 2.76, P = 0.016$). Pupae were first observed on day 9 for the optimal diet and first observed on day 11 for the reduced water diet. More than twofold more larval bees pupated successfully when grafted on the diet based on optimal clover plants’ nutritional make-up compared to those grafted on the diet based on nutritional quality of plants grown under reduced water ($F_{3,5} = 2.51, P = 0.026$). Despite qualitative differences, there was no treatment effect on adult body mass (optimal: 103.6 ± 5.6 mg vs reduced water: 87.1 ± 12.5 mg; $t_{13} = 1.14, P = 0.27$).

Effect of Nutrition on Developing Bumble Bees
Bumble bee microcolonies exhibited higher survival on the optimal diet as compared to the reduced water diet (Kaplan-Meier: $X^2 = 8.6, P = 0.0003$). Notably, foragers lived 60% longer when raised on the optimal diet: 53.9 ± 2.7 d versus 33.6 ± 2.5 d, respectively ($F_{1,76} = 28.634, P < 0.0001$). There was no difference in forager body mass between the treatments (optimal: 91.4 ± 3.8 mg; reduced water: 93.0 ± 4.0 mg). Honey pots were built by 26% of the microcolonies overall, where diet had no effect on whether or not honey pots were present (OR: 1.53, $X^2 = 0.64, P = 0.42$). However, significantly more microcolonies on the optimal diet (13/37) laid eggs than did microcolonies on the lower diet (5/37) ($F_{3,5} = 4.7, P = 0.03$).

We observed similar results for the B. impatiens colonies; colonies raised on the optimal diet had significantly higher individual bee survival than colonies on the reduced diet (Kaplan-Meier:...
Discussion

This is the first study, to our knowledge, to explicitly link water reduction, plant floral resources and bee health via the impact of water on nectar and pollen. We show that the nutritional quality and availability of floral resources for honey bees and bumblebees are significantly decreased by reduced water availability to clover plants. We raised these two commercially important pollinator species on artificial diets determined from the water reduction experiment and found that both bee species experienced significantly lower survival. Bumble bees showed lower incidences of egg laying (microcolonies) and overall productivity (i.e., new bees produced in full colonies). Decreased survival of developing honey bees and of bumble bee foragers may have subsequent colony-level effects. Moreover, a smaller workforce may have cascading negative effects on a social insect colony's ability to collect resources (Schmid-Hempel and Schmid-Hempel 1998), and in the case of honey bees, to produce honey (Woyke 1984). Resource and nutritional intake by bees affect their development and survival. Because of our reliance on bees and bee pollination services (Klein et al. 2007, Kremen et al. 2007), we need to consider how floral resources and the nutrition they provide may shift under mechanisms of global change, such as invasions (Harmon-Threatt and Kremen 2015) and drought (Phillips et al. 2018).

While our artificial diets are able to mimic key aspects of floral nutrition (e.g., relative proportion of sugars, sugar concentration, total protein), it is important to note that these artificial diets were not designed to assess changes in other nutrients. Recent studies have documented that changes in pollen lipids alter the ability of bees to learn (Arien et al. 2015, Arien et al. 2018) as well as their foraging behavior with subsequent impacts on mortality (Vaudo et al. 2016, Moerman et al. 2017). While temperature can influence the sugar concentration and viscosity of royal jelly (Saricaoglu et al. 2019), if and how water availability directly influences royal jelly quality or composition is unknown. Further research is needed to assess how water restriction affects additional nutrients, such as pollen lipids and free amino acids in nectar and the composition of royal jelly.

Despite limitations of an artificial diet, we link the cascading effects of water limitation to pollinator survivorship and productivity via reductions in nectar and pollen quality. Our study demonstrates that plant phenotype, numbers of inflorescences and florets, and floral resource metrics were all affected by a 30% decrease in water. Floral resource phenology and resource availability are known to be factors influencing bee populations (Wray et al. 2014, Dicks et al. 2015) and determining bee fitness (Ogilvie and Forrest 2017). For Osmia lignaria, a solitary bee, increases in resource availability have been found to favor earlier emergence (Farzan and Yang 2018). Similarly, the size of pollen provisions and their quality influences adult body size in the subsocial, small carpenter bee, Ceratina calcarata (Lawson et al. 2016), and adult life span in honey bees (Schmidt et al. 1987, Di Pasquale et al. 2013). Bees, like all insects, are reliant on food provisions to complete development. In some bees, restricted access to food during development can prolong development time (Burkle and Irwin 2009) and decrease gyne production (Goulson et al. 2015, Rotheray et al. 2017). Thus, these experiments can contribute to a proactive approach to managing the indirect impacts of water limitation on bee survivorship and productivity. Because clover plants altered the quality of both pollen and nectar in response to water limitation, we are unable to ascribe lower bee survival and productivity to decreased nutritive value of pollen, nectar, or both. Based on previous studies examining effects of various diets (Schmidt et al. 1987, Rotheray et al. 2017), we currently hypothesize both are critically important. Future experiments could manipulate resources independently and at multiple concentrations to elucidate, which relative combination of resources has the largest effect on solitary and social bee species. While our study provides key insights and links water availability with decreased bee survivorship, further research should examine how wild bees foraging behaviors respond to water-induced changes in floral resources.

It is important to note that the plants in this study were not experiencing drought stress, as defined by Blum and Tuberosa (2018). These were otherwise healthy-looking plants that received...
less than optimum water allotments; despite not being under drought stress, plants in the reduced water treatment took longer to develop, produced fewer flowers and less floral resources, and this ultimately led to lower seed set as compared to the optimal water treatment. These results correspond to other studies that documented phenology shifts (Penuelas et al. 2004) and decreased floral displays (Burkle and Runyon 2016, Phillips et al. 2018) when water is decreased.

Any such decreases in floral abundance will likely affect bee foraging behavior. For example, both bumble bees and honey bees often prefer to forage on larger floral displays (Mitchell et al. 2004, Higginson et al. 2006, respectively). In fact, many insects respond to decreased nutrient quality by increasing resource foraging, including crickets (Srygley and Lorch 2013), ants (Christensen et al. 2010, Dussutour and Simpson 2012), and bees (Wright 1988, Pernal and Currie 2000, Pernal and Currie 2001). Our experimental methodologies prevented such compensatory feeding (or increased consumption of the lower quality food to obtain similar nutritive values) from occurring in honey bees and bumble bee micro colonies. For honey bees, this was achieved by grafting each larva onto a standardized volume of artificial diet. For Bombus microcolonies, this was limited by providing each microcolony with the same amount of food resources. We did not measure total resource collection in the full bumble colonies and thus cannot assess whether compensatory feeding occurred at the colony level. However, we found that bumble bees raised on the reduced water diets fared significantly worse in all metrics measured than bumble bees raised on the optimal diets, despite each colony having unrestricted access to the specified experimental diets. Moreover, most studies looking at compensatory feeding and shifts in foraging behavior have not been conducted under an environmental stressor that is experienced at the landscape level (such as water limitation). Therefore, future work should examine how bees may compensate for landscape-wide declines in nutrition as simulated in our experiments in order to better inform bee management strategies.

The effect of decreased nutrition provided by flowers under reduced water conditions (Phillips et al. 2018) will depend on how water reduction is distributed across the landscape. If a single species or patch experiences reduced water, bees will be able to move on to another patch or more rewarding species. This may not directly impact the bees’ fitness but may have negative impacts on the pollination rates of the plants experiencing reduced water. However, under current predictions of climate change, we anticipate landscape-level water reduction (McDowell et al. 2013). This may take the form of reduced watering in managed landscapes due to regulated water restrictions (Mini et al. 2014, Huntsinger et al. 2017), or of natural drought conditions in natural ecosystems.

We document the cascading effects of limited water and its potential impact on bees. While this study focused on social bee species, solitary taxa are likely similarly affected. Since the size of offspring of solitary bees are determined by the quality and quantity of nectar and pollen in provisioning masses (Stone 1994), any landscape-wide decrease in nutritive value of floral resources will impact solitary bee survivorship and productivity as well. Our findings suggest that widespread reduced water, either due to natural weather patterns or mandatory water restrictions, may have a significant impact on the phenology, floral display, and nectar resources of bee forage. In short, the nutritive quality of flowers will need to be considered when developing strategies for promoting healthy bee populations under future climate conditions.

Data Availability Statement

Data from this study are available from the Dryad Digital Repository: https://doi.org/10.6086/D14X10 (Wilson Rankin 2020).

Supplementary Data

Supplementary data are available at Journal of Insect Science online.

Acknowledgments

We are grateful for the assistance of N. Di with the grafting of honey bee larvae, to Q. McFrederick providing honey bee larva, to J. Hazlehurst and D.T. Rankin for their helpful comments on this manuscript. Bombus colonies were obtained from BioBest USA (RD-EC-031) and kept at UC Riverside under CDFA Permit # 3162. Funding was provided from grants to EWR from the National Honey Board and US Department of Agriculture (Hatch CA-R-ENT-5091-H).

Author Contributions

E.W.R. conceived the study, G.E.L. and S.K.B. conducted the experiments, E.W.R. conducted the statistical analyses, and E.W.R., G.E.L. and S.K.B. contributed to the writing of the manuscript.

References Cited

Alaux, C., F. Duclos, D. Crauser, and Y. Le Conte. 2010. Diet effects on honeybee immunocompetence. Biol. Lett. 6: 562–565.

Andresen, L. C., A. Michelsen, S. Jonasson, I. K. Schmidt, T. N. Mikkelsen, P. Ambus, and C. Beier. 2010. Plant nutrient mobilization in temperate heathland responds to elevated CO2, temperature and drought. Plant Soil. 328: 381–396.

Arien, Y., A. Dag, S. Zarkin, T. Masci, and S. Shafir. 2015. Omega-3 deficiency impairs honey bee learning. Proc. Natl. Acad. Sci. U. S. A. 112: 15761–15766.

Arien, Y., A. Dag, and S. Shafir. 2018. Omega-6:3 ratio more than absolute lipid level in diet affects associative learning in honey bees. Front. Psychol. 9: e1001.

Aupinel, P., D. Fortini, H. Dufour, J. Tasei, B. Michaud, J. Odoux, and M. Pham-Delegue. 2005. Improvement of artificial feeding in a standard in vitro method for rearing Apis mellifera larvae. Bull. Insectology 58: 107.

Bates, D., M. Maechler, and B. Bolker. 2013. lme4: Linear mixed-effects models using S4 classes. http://CRAN.R-project.org/package=lme4

Blum, A., and R. Tuberosa. 2018. Dehydration survival of crop plants and its measurement. J. Exp. Bot. 69: 975–981.

Brunet, J., and M. L. Van Etten. 2019. The response of floral traits associated with pollinator attraction to environmental changes expected under anthropogenic climate change in high-altitude habitats. Int. J. Plant Sci. 180: 954–964.

Brunet, J., and Z. Larson-Rabin. 2012. The response of flowering time to global warming in an alpine plant: the impact of genetics and the environment. Botany 90: 319–326.

Burkle, L. A., and J. B. Runyon. 2016. Drought and leaf herbivory influence floral volatiles and pollinator attraction. Glob. Chang. Biol. 22: 1644–1654.

Burkle, L., and R. Irwin. 2009. The importance of interannual variation and bottom-up nitrogen enrichment for plant-pollinator networks. Oikos. 118: 1816–1829.

California Water Boards. 2016. Urban Water Supplier Conservation Standard for Extended Emergency Regulation Rulemaking. http://www.waterboards.ca.gov/water_issues/programs/conservation_portal/docs/emergency_reg/supplier_standards_effective030116.pdf

Cayan, D. R., E. P. Maurer, M. D. Dettinger, M. Tyree, and K. Hayhoe. 2008. Climate change scenarios for the California region. Clim. Chang. 87: 21–42.
Cnaani, J., R. Schmid-Hempel, and J. O. Schmidt. 2002. Colony development, larval development and worker reproduction in Bombus impatiens Cresson. Insectes Soc. 49: 164–170.

Cnaani, J., J. D. Thomson, and D. R. Papaj. 2006. Flower choice and learning in foraging bumblebees: effects of variation in nectar volume and concentration. Ethology 112: 278–283.

Colman, E. A. 2002. A laboratory procedure for determining the field capacity of soils. Soil Sci. 163: 277–283.

Cruden, R. W., and S. M. Hermann. 1983. Studying nectar? Some observations on the art, pp. 223–241. In B. Bentley and T. Elias (eds.), The biology of nectaries. Columbia University Press, New York, NY.

Descamps, C., M. Quinet, A. Bajot, and A. L. Jacquemart. 2018. Temperature and water stress affect plant-pollinator interactions in Borago officinalis (Boraginaceae). Ecol. Evol. 8: 3443–3456.

Di, N., K. R. Hladun, K. Zhang, T. X. Liu, and J. T. Trumble. 2016. Laboratory bioassays on the impact of cadmium, copper and lead on the development and survival of honeybee (Apis mellifera L.) larvae and foragers. Chemosphere. 152: 530–538.

Di Pasquale, G., M. Salignon, Y. Le Conte, L. P. Belzunces, A. Decourtye, A. Kretzschmar, S. Suchail, J. L. Brunet, and C. Alaux. 2013. Influence of pollen nutrition on honey bee health: do pollen quality and diversity matter? PLoS One 8: e72016.

Dicks, L. V., M. Baude, S. P. Roberts, J. Phillips, M. Green, and C. Carvell. 2015. How much flower-rich habitat is enough for wild pollinators? Answering a key policy question with incomplete knowledge. Ecol. Entomol. 40: 22–35.

Dussutour, A., and S. J. Simpson. 2012. Ant workers die young and colonies collapse when fed a high-protein diet. Proc. Biol. Sci. 279: 2402–2408.

Fang, X., N. C. Turner, G. Yan, F. Li, and K. H. Siddique. 2010. Flower numbers, pod production, pollen viability, and pistil function are reduced and collapse when fed a high-protein diet. Proc. Biol. Sci. 279: 2402–2408.

Farzan, S., and L. H. Yang. 2018. Experimental shifts in phenotype affect fitness, foraging, and parasitism in a native solitary bee. Ecology 99: 2187–2195.

Gallagher, M. K., and D. R. Campbell. 2017. Shifts in water availability mediate plant-pollinator interactions. New Phytol. 215: 792–802.

Gérard, M., M. Vanderplanck, T. Wood, and D. Michez. 2020. Global warming and plant–pollinator mismatches. Emerg. Top. Life Sci. 4: 77–86.

Goulson, D., E. Nicholas, C. Botsias, and E. L. Rotheray. 2015. Bee declines driven by combined stress from parasites, pesticides, and lack of flowers. Science. 347: e1255957.

Harde, L. D. 1986. Effects of nectar concentration and flower depth on flower handling efficiency of bumble bees. Oecologia. 69: 309–315.

Harmon-Threatt, A. N., and C. Kremen. 2015. Bumble bees selectively use native and exotic species to maintain nutritional intake across highly variable and invaded local floral resource pools. Ecol. Entomol. 40: 471–478.

Higginsson, A. D., F. S. Gilbert, and C. J. Barnard. 2006. Morphological correlates of nectar production used by honeybees. Ecol. Entomol. 31: 269–276.

Huntstinger, L., T. V. Hruska, J. L. Oviedo, M. W. K. Shapero, G. A. Nader, R. S. Ingram, and S. R. Reissinger. 2017. Save water or save wildlife? Water use and conservation in the central Sierran foothill oak woodlands of California, USA. Ecol. Soc. 22: e12.

Intergovernmental Panel on Climate Change. 2002. Climate change and biodiversity. IPCC Technical Paper 5: 1–86.

James, R., and T. L. Pitts-Singer (eds.). 2008. Bee pollination in agricultural ecosystems. Oxford University Press, New York, NY.

Jordano, P., J. Bascompte, and J. M. Olesen. 2006. The ecological consequences of complex topology and nested structure in pollination webs, pp. 173–199. In N. M. Waser and J. Ollerton (eds.), Plant-Pollinator Interactions: From Specialization to Generalization. University of Chicago Press, Chicago, IL.

Kearn, C. A., and D. W. Inouye. 1993. Techniques for pollination biologists. University Press of Colorado, Niwot, CO.
Schmidt, J. O., S. C. Thoenes, and M. D. Levin. 1987. Survival of honey-bees, *Apis mellifera* (Hymenoptera, Apidae), fed various pollen sources. Ann. Entomol. Soc. Am. 80: 176–183.

Smart, M. D., C. R. V. Otto, and J. G. Lundgren. 2019. Nutritional status of honey bee (*Apis mellifera L.*) workers across an agricultural land-use gradient. Sci. Rep. 9: e16232.

Srygley, R. B., and P. D. Lorch. 2013. Coping with uncertainty: nutrient deficiencies motivate insect migration at a cost to immunity. Integr. Comp. Biol. 53: 1002–1013.

Stone, G. N. 1994. Activity patterns of females of the solitary bee *Anthophora plumipes* in relation to temperature, nectar supplies and body-size. Ecol. Entomol. 19: 177–189.

Therneau, T. M. 2015. A Package for Survival Analysis in S. version 2.38. https://CRAN.R-project.org/package=survival.

Vanderplanck, M., B. Leroy, B. Wathelet, R. Wattiez, and D. Michez. 2014. Standardized protocol to evaluate pollen polypeptides as bee food source. Apidologie. 45: 192–204.

Vaudo, A. D., D. Stabler, H. M. Patch, J. F. Tooker, C. M. Grozinger, and G. A. Wright. 2016. Bumble bees regulate their intake of essential protein and lipid pollen macronutrients. J. Exp. Biol. 219: 3962–3970.

Wang, W., R. Vignani, M. Scali, and M. Cresti. 2006. A universal and rapid protocol for protein extraction from recalcitrant plant tissues for proteomic analysis. Electrophoresis. 27: 2782–2786.

Waser, N. M., and M. V. Price. 2016. Drought, pollen and nectar availability, and pollination success. Ecology 97: 1400–1409.

Wilson Rankin, E. E., S. K. Barney, and G. E. Lozano. 2020. Data from: reduced water negatively impacts social bee survival and productivity via shifts in floral nutrition. Dryad Digital Repository. doi:10.6086/D14X10.

Winfree, R. 2010. The conservation and restoration of wild bees. Ann. N. Y. Acad. Sci. 1195: 169–197.

Woyke, J. 1984. Correlations and interactions between population, length of worker life and honey production by honeybees in a temperate region. J. Apic. Res. 23: 148–156.

Wray, J. C., L. A. Neame, and E. Elle. 2014. Floral resources, body size, and surrounding landscape influence bee community assemblages in oak-savannah fragments. Ecol. Entomol. 39: 83–93.

Wright, D. H. 1988. Temporal changes in nectar availability and *Bombus appositus* (Hymenoptera: Apidae) foraging profits. Southwest. Nat. 33: 219–227.