Size-dependent influence of NOx on the growth rates of organic aerosol particles

C. Yan1,2*, W. Nie3,4*, A. L. Vogel5,6, L. Dada3, K. Lehtipalo1,4,5*, D. Stolzenburg6, R. Wagner1, M. P. Rissanen1, M. Xiao2, L. Ahonen1, L. Fischer7, C. Rose16, F. Bianchi1,8, H. Gordon3,9‡, M. Simon10, M. Heinritzi10, O. Garmash1, P. Rollin11, A. Dias3,12, P. Ye13,14, V. Hofbauer13, A. Amorim13, P. S. Bauer12, A. Bergen10, A.-K. Bernhammer12, M. Breitenlechner7, S. Brilke4,10, A. Buchholz5, S. Buenrostro Mazon1, M. R. Canagaratna3, X. Chen3,15, A. Ding4, J. Dommen7, D. Craper16, J. Duplissy1, C. Frege1, C. Heyn9, R. Guida2, J. Hakala1, L. Heikkilä1, C. R. Hoyle4,5, T. Jokinen1, J. Kangaslampi1,8, J. Kirkby10, J. Kontkanen1, A. Kürten10, M. J. Lawler16, H. Mai17, S. Mathot3, R. L. Mauldin III13,18, U. Molteni4, L. Nickovic19,‡, T. Nieminen1, J. Nowak10,‡, A. Ojdanic6, A. Onnela3, A. Pajunoja15, T. Petäjä1,2, F. Piol10,§, L. L. J. Quéléver1, N. Sarnela1, S. Schallhart1, K. Sengupta4, M. Sipila3,1, A. Tome7, J. Tröstl4, O. Viisanen10, A. C. Wagner10,9, A. Yli-Sirén15, Q. Zha3, U. Baltensperger4, K. S. Carslaw9, J. Curtius10, R. C. Flagan17, A. Hansel1,7,21, I. Riipinen22, J. N. Smith16, A. Virtanen15, P. M. Winkler6, N. M. Donahue13, V.-M. Kerminen1, M. Kulmala1,2,8,23, M. Ehn1, D. R. Worsnop1,14,15

Atmospheric new-particle formation (NPF) affects climate by contributing to a large fraction of the cloud condensation nuclei (CCN). Highly oxygenated organic molecules (HOMs) drive the early particle growth and therefore substantially influence the survival of newly formed particles to CCN. Nitrogen oxide (NOx) is known to suppress the NPF driven by HOMs, but the underlying mechanism remains largely unclear. Here, we examine the response of particle growth to the changes of HOM formation caused by NOx. We show that NOx suppresses particle growth in general, but the suppression is rather nonuniform and size dependent, which can be quantitatively explained by the shifted HOM volatility after adding NOx. By illustrating how NOx affects the early growth of new particles, a critical step of CCN formation, our results help provide a refined assessment of the potential climatic effects caused by the diverse changes of NOx level in forest regions around the globe.

INTRODUCTION

Atmospheric new-particle formation (NPF) contributes to about half of the global tropospheric cloud condensation nuclei (CCN) population (1), thereby affecting Earth’s radiation balance via aerosol-cloud interactions (2). However, considerable uncertainties exist on how atmospheric NPF and CCN production are associated with anthropogenic emissions of different aerosol precursor gases. The main reasons for these uncertainties are our incomplete knowledge on the mechanisms that dictate NPF and subsequent growth of newly formed particles to CCN sizes in the atmosphere.

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1Institute for Atmospheric and Earth System Research/NAR-Physics, Faculty of Science, University of Helsinki, 00560 Helsinki, Finland. 2Joint International Research Laboratory of Atmospheric and Earth System Sciences; School of Atmospheric Sciences, Nanjing University, Nanjing, China. 3CERN, CH-1211, Geneva, Switzerland. 4Laboratory of Atmospheric Chemistry, Paul Scherrer Institute, 5232 Villigen, Switzerland. 5Finnish Meteorological Institute, Eri Palkomäki aukio 1, 00560 Helsinki, Finland. 6University of Vienna, Faculty of Physics, Boltzmanngasse 5, 1090 Wien, Austria. 7University of Innsbruck, Institute for Ion and Applied Physics, 6020 Innsbruck, Austria. 8Aerosol and Haze Laboratory, Beijing Advanced Innovation Center for Soft Matter Science and Engineering, Beijing University of Chemical Technology, Beijing, China. 9Division of Nuclear Physics, Department of Physics, Lund University, P. O. Box 118, SE-221 00 Lund, Sweden. 10CENTA and FCUL, Universidade de Lisboa, Campo Grande, 1749-016 Lisboa, Portugal. 11Carnegie Mellon University Center for Atmospheric Particle Studies, 5000 Forbes Ave, Pittsburgh, PA 15213, USA. 12Aerodyne Research Inc., Billerica, MA 01821, USA. 13University of Eastern Finland, Department of Applied Physics, P.O. Box 1627, 70211 Kuopio, Finland. 14Department of Chemistry, University of California, Irvine, CA 92697, USA. 15Centra Institute of Technology, 210-41, Pasadena, CA 91125, USA. 16Department of Chemistry and Biochemistry, University of Colorado, Boulder, CO 80309, USA. 17School of Earth and Environmental Science, University of Manchester, Manchester M13 9PL, UK. 18Department of Environmental Science and Analytical Chemistry (ACES) and Bolin Centre for Climate Research, Stockholm University, 10691 Stockholm, Sweden. 19Institute of Physics, University of Tartu, FI-00014 Helsinki, Finland.

*These authors contributed equally to this work.

†Corresponding author. Email: chao.yan@helsinki.fi (C.Y.); katrianne.lehtipalo@helsinki.fi (KL)

‡Present address: Atmospheric Composition Research, Finnish Meteorological Institute, 00101 Helsinki, Finland.

§‡‡Present address: Department of Chemistry, University of Oslo, 0315 Oslo, Norway.

¶‡‡Present address: Department of Chemistry, University of Tartu, FI-00014 Helsinki, Finland.

#Present address: Institute for Atmospheric and Climate Science, ETH Zurich, Switzerland.

††Present address: Department of Chemistry, University of Colorado, Boulder, CO 80309-0215, USA.

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NPF consists of two consecutive steps: particle nucleation forming small clusters (usually 1 to 2 nm) and their further growth to larger sizes (3). The efficiencies of both steps together determine the rate of CCN formation from NPF. Particle nucleation produces an initial pool of newly formed particles, and these particles need to grow sufficiently fast to avoid being scavenged by the large preexisting particles (4). Organic vapors play crucial roles in both steps of NPF. Under most tropospheric conditions, particle nucleation is prevailingly driven by sulfuric acid (3, 5), but organic vapors might act as an important stabilizing agent of sulfuric acid clusters (6, 7). Organic vapors dominate particle growth in most tropospheric conditions (8, 9) and therefore are crucial for the survival of newly formed particles.

The role of organic vapors in NPF differs significantly according to the volatility, which can span over 10 orders of magnitude (10). The formation and survival of newly formed particles are responsive to only a small fraction of organic vapors, which have (extremely) low volatility and thus are capable of clustering with themselves or sulfuric acid (11, 12), and more readily condense onto the smallest particles and favor their survival from scavenging loss (13–15). Although observational evidence has suggested the existence of such low-volatility organic vapors (15), their identity and sources have been a puzzle for many years. Very recently, the autoxidation of peroxy radicals (RO₂), involving a few steps of intramolecular H migration and subsequent O₂ addition (16), has been found as the most efficient pathway of forming these low-volatility vapors (17). Chemically, these vapors are highly oxygenated and therefore are also widely referred to as highly oxygenated organic molecules (HOMs) (18).

Nitrogen oxides (NOₓ), mainly emitted by anthropogenic activities nowadays, are key players in atmospheric chemistry through their reactions with other radicals (19). Their role in regulating atmospheric oxidants is well established (20), with a direct impact on volatile organic compound (VOC) oxidation processes and consequently on the formation of condensable organic vapors. NOₓ has been found to significantly suppress NPF from monoterpene oxidation (22, 23), although the cause was only speculated, lacking direct observations on a molecular level. As HOMs have been known as key precursors in NPF in multiple studies, it is foreseeable that investigating the influence of NOₓ on them can help understand the details about this “NOₓ suppression of particle formation.” Two recent papers have addressed the role of NOₓ on atmospheric autoxidation, suggesting that reduced NOₓ concentrations will make autoxidation increasingly more important in the future (24), although the HOM formation rates from increased autoxidation may be counteracted by a concurrent decrease in oxidant concentrations (25). However, while these studies have addressed an overall “bulk” HOM formation potential, NOₓ will affect not only the total HOM yield but also their composition and, thereby, their physical properties. Currently, our understanding of the effect of NOₓ on particle formation can be improved from two aspects. First, the changes caused by NOₓ on the chemical composition and bulk volatility of HOMs need to be understood based on direct measurement on a molecular level; second, the response of particle formation and growth to those changes in HOMs needs to be investigated in detail.

We conducted well-controlled NPF experiments using the CLOUD (Cosmics Leaving Outdoor Droplets) chamber equipped with a collection of state-of-the-art instruments. The comprehensive measurements allowed for obtaining important details of the NPF, from formation of low-volatility vapors to particle nucleation and further growth. After adding NOₓ at levels up to only a few parts per billion by volume (ppbv), we observed substantial changes in both the growth rates (GRs) of new particles and HOM composition. We performed detailed analysis on HOM volatility based on their thermal desorption temperature and found that NOₓ led to a significant increase of HOM volatilities, which, in turn, could be quantitatively connected to the suppression of particle GRs in a size-dependent manner.

RESULTS

We performed a set of experiments in the CLOUD chamber at CERN to investigate the effect of NOₓ on particle formation via modifying the HOM composition. A typical experiment started with an injection of ozone and monoterpines without any NOₓ. A mixture of α-pinene and Δ-3-carene with 2:1 volume mixing ratio was used as VOC precursors to better resemble the monoterpene profile in a boreal forest station [SMEAR II (Station for Measuring Ecosystem-Atmosphere Relations)] in southern Finland (26). The ultraviolet (UV) system was kept on throughout the experiment which produces hydroxyl radicals (OH) by photolyzing O₃ (see Materials and Methods), in addition to those OH from ozonolysis of monoterpenes. These two sources together resulted in a steady-state OH concentration of around 10⁶ cm⁻³. Similar to previous CLOUD experiments (12), we started a typical experiment with adding monoterpene and ozone with zero ions in the chamber (referred to as neutral condition) and a relatively weak NPF occurred. We then turned off the high voltage [referred to as galactic cosmic ray (GCR) condition] and allowed the ions to trigger a stronger NPF that is distinguishable from the weaker one (see Fig. 1B). After the nucleation rate stabilized and particles grew to a few tens of nanometers, we injected NO into the chamber, which was oxidized mostly to NO₃ by O₃ (NO:NO₃ about 1%) and a small fraction further to NOₓ (NO₃:NOₓ about 0.007%). As the NPF became progressively weaker when NOₓ level increased, it is impossible to separate the subsequent weaker NPF case from the former stronger one. Therefore, when increasing the NOₓ level, we also removed all ions for roughly 15 min (by turning on the high voltage) to quench the former nucleation, which resulted in a clear “gap” between the NPF cases (see Fig. 1B). The change of the HOM composition in both gas and particle phases as well as the particle dynamics were measured at three different NOₓ concentration levels. Experiments with similar procedure were conducted with different monoterpene and SO₂ concentrations, except for experiment 1748, in which the “NPF gaps” were not inserted for experiment 1748, in which the “NPF gaps” were not inserted between NPF events, causing later events to be undistinguishable (see table S1). Other detailed information about the chamber operation and experimental conditions can be found in Materials and Methods.

NPF at different NOₓ levels

We show in Fig. 1 one example of the resulting data (experiment 1752, see table S1), demonstrating the influence of NOₓ on the HOM formation and NPF. A stepwise increase in the NOₓ concentration caused an evident change in the HOM composition, featured by a large increase of the HOM monomer (4 ≤ carbon number ≤ 10) concentration and a simultaneous decrease of the HOM dimer (10 < carbon number ≤ 20) concentration at higher NOₓ concentrations (Fig. 1A). The monoterpene concentration did not change notably during the addition of NOₓ, so the amount of monoterpenes available
H2 SO4), and the large reduction of experiments (see fig. S1 and table S1). The suppression degrees are only valid for this specific condition and will vary in other cases, and the ratios represent the suppression by NOx. It should be noted that such changes in the concentrations of oxidants were observed, but these changes were relatively small and cannot explain the observed major change in the HOM composition.

New particles were being formed continuously throughout the course of the experiment, where stronger bursts of new particles were observed for GCR conditions (see Supplementary Materials). We show in Fig. 1B that separated NPF events were triggered at different NOx levels. We characterize these NPF events by their particle formation rate at 1.7 nm (J1,7) and GRs in different size ranges (see the Supplementary Materials). As shown in our previous work, the nucleation in these experiments was driven by biogenic vapors (independent of H2SO4), and the large reduction of J1,7 at higher NOx levels (Fig. 1C) is a result of the decreased HOM dimer concentration \( (11) \).

We calculated the GRs of newly formed particles for different size ranges according to their appearance time into these sizes (see the Supplementary Materials). We found that the baseline GRs, measured at the zero NOx condition, were 8.1 ± 4.5, 12.2 ± 4.7, 20.2 ± 3.7, 22.0 ± 0.9, 14.1 ± 0.6, and 10.6 ± 0.2 nm/hour in the size ranges of 1.3 to 2.3, 1.9 to 3.5, 3.5 to 7, 7 to 20, 20 to 30, and >30 nm, respectively. Size-segregated GRs after adding NOx were also calculated and normalized to these baseline values (Fig. 1D). We found that the GRs at different size ranges were reduced by different degrees: The suppression was most pronounced for the smallest particles and increasingly weaker for larger ones, eventually becoming almost negligible for particles larger than 30 nm in diameter. These findings clearly indicate that the effect of NOx cannot be thought of simply as an overall suppression on the full course of NPF, since otherwise, the GR at each size interval should have changed in the same way. Instead, the results suggest a more complicated change in the volatility distribution of condensable vapors.

We list size-segregated particle GRs determined for all experiments in table S1 and plot the normalized GRs in fig. S1. Experiments 1749 to 1752 were conducted with a high level of monoterpenes, and the strongest NPF events were observed in these experiments—the events were still distinguishable even when moderate-level NOx was injected. However, reliable determination of particle GRs is challenging for experiments with lower monoterpene concentrations, e.g., experiments 1753 to 1755, when the NPF events were much weaker. The presence of H2SO4 led to less pronounced reduction of GRs, as the contribution of H2SO4 to particle growth was almost unaffected by NOx. This can be seen by comparing experiments with similar monoterpane concentration but different H2SO4 concentration, e.g., experiments 1749 and 1752. We also noticed that there are some likely increase of \( GR_{7-20 \text{ nm}} \) and \( GR_{20-30 \text{ nm}} \) along with the increase of NOx when H2SO4 is present (e.g., experiments 1749 and 1750; fig. S1). This indicates that H2SO4 may interact with HOMs on the surface of big particles, leading to an enhancement of HOM condensation. However, since the increase of GRs is not prominent (around 20% maximum) and is within the uncertainty range, it is difficult to fully validate this interpretation. For accuracy reasons, we use experiment 1752 as the best example to show the effect of NOx, but overall, the size-dependent suppression on particle growth is evident in all experiments as long as the GRs can be well determined.

HOM composition modified by NOx

We next investigate molecular-level changes in the HOM composition between conditions with and without NOx, measured with a chemical ionization atmospheric-pressure-interface time-of-flight mass spectrometer (CI-API-TOF, see Supplementary Materials). Changes in the HOM composition were observed immediately after NOx was injected and were sensitive to the change of NOx concentration (fig. S2). Here, we show the changes in HOM composition when 1.9 ppbv NOx was added in comparison to that without NOx. This condition is chosen as it better represents the typical NOx level (about 1.5 ppbv) and also best resembles the HOM composition observed in a boreal forest station (SMEAR II) (fig. S2). To better describe the behavior of HOMs, in addition to the division of HOM monomers (marked with a subscript “mono”) and dimers (subscript “di”), we further group them according to the number of contained nitrogen atoms (0, 1, or 2), which are marked with \( C_{x}H_{y}O_{z}N_{n} \) and \( C_{x}H_{y}O_{z}N \), respectively.

Before the injection of NOx, the HOMs formed in our experiment were mostly \( C_{6}H_{10}O_{2,mono} \) (78.9%) and \( C_{6}H_{10}O_{2,di} \) (19.4%) compounds with a tiny fraction (1.7%) of residual \( C_{6}H_{10}O_{2,N_{mono}} \) compounds.
left from the previous experiment (Fig. 2, A and C). The presence of 1.9 ppbv NOx resulted in the formation of organic nitrates (Fig. 2, B and D), including C19H28O9N mono (25.4%), C20H31O8N mono (3.0%), C19H28O9N di (3.6%), and C20H31O8N2 di (0.5%). Meanwhile, C19H28O9 mono and C19H28O9 di decreased to 60.7 and 6.7%, respectively. The evolution of these species in the full course of the experiment can be seen in fig. S3.

The HOM formation is a result of several complicated reactions, in which the reactions between NOx and RO2 play an important role (19, 27). We are not aiming to determine the exact contributions of all reaction pathways here, so instead, we summarize below the most important aspects of HOM formation and provide the supporting observational evidence in the Supplementary Materials. First, we found that the presence of NOx had a small impact on the overall compositional change of gas-phase HOMs. The pie charts give the fractional contribution of different types of HOMs. (C and D) The spectra of HOMs colored by their types. The x axis is the exact mass of HOMs, and the y axis is the mass defect. The color of circles denotes the type of HOMs, and their size is proportional to the logarithm of the count rate. Each straight line represents a group of compounds with the same number of carbon, hydrogen, and nitrogen but different numbers of oxygen atoms. The line style is the same as that used for the annotation frame.

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In addition to the gas-phase HOMs, we also measured the particle-phase HOMs using the filter inlet for gases and aerosols (FIGAERO) coupled to an iodide-based chemical ionization time-of-flight mass spectrometer (see Supplementary Materials). The changes in the particle-phase HOM composition were generally similar to that of the gas-phase HOMs, featured by the increase of all types of organic nitrates and the decrease of non-nitrate HOMs, especially dimers (C19H28O9 di) (fig. S2). The simultaneous change of HOM composition in both gas and particle phases is a strong evidence that the particle formation is directly affected by condensation of gas-phase HOMs.

Change of HOM volatility distribution by NOx

We finally estimate how the altered HOM composition changes the HOM volatility, a key parameter that governs HOM condensation and, therefore, connects the HOM chemistry with the particle growth behavior. We first investigated the HOM volatilities according to their thermal-desorption temperature (Tmax see the Supplementary Materials). Figure 3A shows the thermograms of three representative dimer compounds at different NOx levels, i.e., C19H28O9 (400.17 Th), C20H31O8N (413.20 Th), and C20H31O8N2 (476.20 Th), representing C19H28O9 mono, C20H31O8N di, and C20H31O8N2 di compounds. Consistent with former observations of HOMs in the gas phase, along with the increase of NOx, C19H28O9 decreases in contrast to the increase of C20H31O8N and C20H31O8N2.

Although C20H31O8N2 has a higher molecular weight and a larger oxygen-to-carbon ratio (O:C) than C19H28O9, it desorbs at a lower temperature, suggesting a higher volatility. Although the Tmax of HOM dimers shows a weak dependence on the molecular weight (Fig. 3B), the large discrepancy of Tmax between different types suggests that the molecular weight is not the most crucial parameter for their volatility. We found that Tmax is strongly correlated with the effective O:C (O:Ceff) regardless of the HOM dimer type (Fig. 3C), indicating that O:Ceff can be a good reference to their volatility. Here, the O:Ceff is calculated based on the directly measured carbon and oxygen numbers but subtracting two oxygen atoms for each

![Fig. 2. Gas-phase HOMs under zero and 1.9 ppbv NOx conditions measured by CI-API-TOF in the CLOUD chamber. (A and B) The spectra of HOMs colored by their types. The pie charts give the fractional contribution of different types of HOMs. (C and D) Mass defect plots showing HOM composition under the two conditions.](http://advances.sciencemag.org/)
Fig. 3. Thermal desorption of particle-phase HOM dimers measured with the FIGAERO. (A) The thermogram of three example molecules under different NOx conditions. Different line styles represent different NOx conditions. The $T_{\text{max}}$ is defined as the temperature at which the signal intensity reaches the maximum. (B) Correlation between $T_{\text{max}}$ and mass-to-charge ratio for all HOM dimers. (C) Correlation between $T_{\text{max}}$ and the effective O:C for all HOM dimers. The size of the circles in (B) and (C) is linearly proportional to the signal intensity of the desorption thermogram.

nitrogen atom. The use of $O:C_{\text{eff}}$ can be justified by the volatility dependence on the functional groups: It is suggested that the alcohol (-OH) and nitrate (-ONO$_2$) groups, contributing the same to $T_{\text{max}}$, should hold for monomers as well. Therefore, the relationship between $T_{\text{max}}$ and volatility of monomers should hold for monomers as well. However, the fundamental assumption that the volatility is affected by functional groups should hold for monomers as well. Therefore, the relationship between $T_{\text{max}}$ and volatility of monomers is expected to follow the same behavior. While we conclude that the influence of molecular weight on HOM volatility is minor when comparing different HOMs with similar carbon numbers, it becomes important for molecules containing very different numbers of carbon atoms, such as a HOM monomer versus a HOM dimer.

In our previous work, we parameterized the HOM volatility to the apparent $O:C$ (14), which is equal to the $O:C_{\text{eff}}$ for HOMs not containing nitrogen atoms. Here, we show that the same parameterization can be extended to other types of HOMs by simply replacing the apparent $O:C$ with the $O:C_{\text{eff}}$. The volatility of all types of HOMs in this study is thus estimated as $\log_{10}C_{\text{eff}} = (0.672 - O:C_{\text{eff}})/0.078$ and $\log_{10}C_{\text{max}} = (0.209 - O:C_{\text{eff}})/0.052$ for HOM monomers and dimers, respectively.

After applying the volatility parameterization to all HOMs, we can obtain the overall HOM volatility distribution by grouping them into volatility bins. For simplicity, we compare the volatility distribution at zero NOx and 1.9 ppbv NOx (Fig. 4A). In both cases, the volatility spans a large range from extremely low-volatility organic compounds (ELVOCs, $C^* \leq 10^{-4.5} \mu g m^{-3}$ or roughly equivalent to $N^* \leq 5 \times 10^4$ cm$^{-3}$, assuming an average molar mass of 300 Da), through low-volatility organic compounds (LVOCs, $10^{-4.5} < C^* \leq 10^{-3.5} \mu g m^{-3}$; $5 \times 10^4 < N^* \leq 5 \times 10^5$ cm$^{-3}$), and on to semivolatile organic compounds (SVOCs, $10^{-3.5} < C^* \leq 10^{-2.5} \mu g m^{-3}$ or $5 \times 10^5 < N^* \leq 5 \times 10^{11}$ cm$^{-3}$). Adding NOx considerably shifts the overall distribution toward a higher volatility. As shown in fig. S5, the fractional decrease of ELVOCs is mostly due to the suppressed $C_{\text{H}_2}\text{O}_x$ di formation by NOx, which is slightly compensated by the formation of $C_{\text{H}_2}\text{O}_x\text{N}_{1-2}$ di. The increase of LVOCs and SVOCs mostly results from the formation of $C_{\text{H}_2}\text{O}_x\text{N}_{1-2}$ mono. Simply put, NOx suppresses dimer formation and replaces dimers with organic nitrate monomers.

Since the particle GR depends approximately linearly on the concentration of condensable vapors, the ratio of particle GRs at different NOx conditions shown in Fig. 1 should be reflective of the corresponding HOM concentration ratios. Figure 4B shows the cumulative HOM concentrations for the two NOx conditions together with their ratio, which increases from ~0.2 for the nonvolatile HOMs ($C^* \leq 10^{-15} \mu g m^{-3}$ or $N^* \ll 1$ cm$^{-3}$) close to unity when counting all LVOCs ($C^* \leq 10^{-0.5} \mu g m^{-3}$ or $5 \times 10^6$ cm$^{-3}$). This volatility measurement, relying on the identity and $T_{\text{max}}$ of HOMs, provides the confirmation of our initial hypothesis that the change in HOM composition caused by NOx is indeed able to explain the observed size-dependent GRs: The abundance of the least volatile vapors is substantially reduced, thus affecting the growth of the smallest particles, while the total concentration of vapors able to condense onto particles with diameters of a few tens of nanometers remains similar.

The tight connection between HOM volatility and particle formation is also supported by the correlogram between the cumulative HOM concentration and particle formation and GRs using data from all experiments listed in table S1. As shown in fig. S6, the correlation coefficient for $GR_{1.9-3.5 \text { nm}}$ is at maximum for HOMs with $C^* \leq 10^{-2.5} \mu g m^{-3}$ ($N^* \leq 5 \times 10^5$ cm$^{-3}$), and quickly decreases if HOMs of higher volatility are included, indicating that these higher-volatility HOMs do not contribute to the particle growth in this size range. Such quick decline of the correlation coefficient for the $GR_{10-30 \text { nm}}$ only occurs when HOMs of $C^* \geq 10^{-2.5} \mu g m^{-3}$ ($N^* \geq 5 \times 10^6$ cm$^{-3}$) are counted, showing a less strict volatility requirement for growing 20- to 30-nm particles owing to the diminished curvature effect. Moreover, the correlations for $J_{1.7}$ and $GR_{1.9-3.5 \text { nm}}$ show very similar patterns, suggesting that the formation and growth of particles at these size ranges are likely led by the same vapors or at least by vapors with nearly fixed relative yields.

DISCUSSION

In summary, using the CERN CLOUD facilities, we performed dedicated experiments to investigate the role of NOx in the particle growth under conditions that mimic the atmosphere in a boreal forest.
and attributed this to the reduced HOM dimer formation and suppressed particle growth can be quantitatively connected to the suppression effect of NOx on particle formation. The comprehensive measurements of both the particle precursors vapors (HOMs) and the growth over various size ranges. Our results show a generally consistent picture with the few recent studies that NPF is suppressed by NOx due to the change of HOM chemistry. However, on the basis of the detailed analysis on particle dynamics, we reveal that the suppression effect of NOx on particle formation is rather nonuniform and size dependent. Furthermore, we elucidate that the size-dependent suppression on particle growth can be quantitatively connected to the increased HOM bulk volatility as a result of changes in the HOM chemistry and composition.

In our experiments, we also observed suppressed particle formation rates \( f_i \) and attributed this to the reduced HOM dimer formation \( \left( f_{11} \right) \). However, this effect becomes significantly weaker when also NH3 and H2SO4 are present \( \left( f_{11} \right) \). This means that the suppression effect of NOx on particle nucleation, reported in previous studies \( \left( f_{11} \right) \) and in this work, cannot be directly applied to the atmosphere where NH3 or even stronger bases together with H2SO4 tend to drive the nucleation.

However, unlike the particle nucleation, the suppression of particle growth driven by HOMs is directly relevant to the ambient atmosphere. After adding NOx, we observed much stronger influence (suppression) on particle growth of small particles (~2 nm), while that on large particles (>30 nm) was negligible. This observation has important implications. First, as smaller particles are more easily scavenged by preexisting particles, the attenuated GR of small particles significantly reduces the survival probability of the newly formed particles and thus causes a reduction of the concentration of CCN-size particles, as seen in our experiments. Second, it also provides a plausible explanation for the laboratory observations showing that NOx has a smaller effect on the yield regarding secondary organic aerosol (SOA) formation than on NPF \( \left( f_{22}, f_{23} \right) \). In addition, our results show that NO is more effective than NO2 in changing the HOM composition and volatility. This indicates that, besides the commonly used term “VOC/NOx,” the NO:NO2 ratio is another crucial parameter in understanding the influence of NOx on SOA formation.

Fig. 4. Volatility distribution of gas-phase HOMs under zero and 1.9 ppbv NOx conditions. (A) The summed HOM concentrations of each bin. (B) The cumulative HOM concentrations. Red and blue markers denote HOM concentrations under zero and 1.9 ppbv NOx, respectively. The black dots give the ratio of cumulative concentrations of HOM\(_{\text{NOx}}\)/HOM\(_{\text{v/o NOx}}\).

with slight influence by human activities. The comprehensive measurements of both the particle precursors vapors (HOMs) and the particle dynamics allow us to evaluate the influence of NOx at all stages of gas-to-particle conversion—from the oxidation of VOC forming HOMs to the particle nucleation and the subsequent particle growth. Our results suggest that in monoterpene-rich environments, such as forested areas, NOx can significantly reduce the CCN formation and thereby influence cloud properties. Our experimental insights, as presented above, can also help improve the modeling of such effects.

MATERIALS AND METHODS

The CLOUD facility

The CLOUD chamber is a stainless steel cylinder with a volume of ca. 26.1 m3, located at CERN, Geneva, Switzerland. The most important feature of this chamber is its ultracleanliness, which allows one to study the NPF phenomenon under carefully controlled and atmospherically relevant conditions, i.e., with precursors of similar concentrations to those in the atmosphere. Dedicated efforts are made to ensure a low contamination level in the chamber; besides the electropolished inner surfaces of the chamber, vigorous rinsing with ultrapure water at 373 K is done before each campaign, and ultraclean synthetic air produced by mixing cryogenic liquid nitrogen and oxygen is used throughout the experiments. The background total VOC concentration is at sub-ppbv level, and the total condensable vapor concentration is at sub-pptv (parts per trillion by volume) level. Ion concentrations in the chamber can be controlled with a high-voltage clearing field. By turning on the high-voltage field (20 kV m\(^{-1}\)), all ions and charged particles are removed; we refer to this as the neutral condition state. When the high voltage is switched off, ions are produced by the GCR in the chamber; we refer to this as the GCR condition.

To mimic the photochemistry caused by sunlight in the atmosphere, a UV light system was used. The system consists of three light sources that cover different regions of the UV and visible spectrum. A krypton fluoride excimer UV laser (3 W, \( \lambda = 248 \) nm) is used to produce OH via O3 photolysis. Two UV light-emitting diodes (LEDs; 2 × 16.5 W, \( \lambda = 370 \) to 390 nm) are used to photolyze NO2 into NO. In addition, four Hamamatsu Xenon arc lamps (4 × 200 W, \( \lambda = 250 \) to 580 nm) are used to provide broad range UV light and bring the overall UV spectrum closer to atmospheric levels. The xenon arc light and the UV laser are fed vertically through the top of the chamber by optical fibers, while the UV LEDs shine into the chamber horizontally from opposite sides in the middle plane. All gases are injected through a dedicated inlet system from the bottom of the chamber. In order to improve the homogeneity of gas mixing inside the big chamber volume, two mixing fans are mounted on the top and the bottom of the chamber.
Experimental design
We conducted a set of experiments for studying the effect of NO$_3$ on HOM production and NPF at constant temperature (278 K) and relative humidity (38%). We kept the injection rate of ozone constant throughout the experiments, thereby maintaining an ozone volume mixing ratio of ca. 40 ppbv. We started a typical experiment with adding monoterpenes under neutral conditions. The monoterpen precursors were a mix of α-pinene and Δ3-carene, the two most abundant monoterpenes at the Hyytiälä station with an initial mixing ratio of 2:1 (26); these compounds are structurally similar, both having one endocyclic double bond on the six-carbon ring. Once the HOM concentration reached steady state and the nucleation rate also stabilized, we turned off the high voltage and allowed the ion concentration to build up, which is referred to in CLOUD experiments as the GCR condition. The ions triggered a stronger particle nucleation that can be easily distinguished from the previous, weaker one under neutral conditions. After the nucleation rate at the GCR condition reached the plateau and the particles grew to a particle nucleation that can be easily distinguished from the previous, order of magnitude lower than that at the Hyytiälä station in spring season. After all types of HOMs reached steady state and a stable run, which was repeated with various initial monoterpene concentrations coupled with different initial SO$_2$ concentrations. Throughout the run, the UV light system was kept on to avoid any change in NPF associated to a varied UV irradiation. The main experimental variables are listed in table S1.

We monitored the NPF events with a variety of instruments (see the Supplementary Materials) and calculated size-resolved particle GRs according to their appearance time (see the Supplementary Materials and fig.S6). In addition, we deployed two chemical ionization mass spectrometers (CIMS) to extend observations of NPF into a molecular level: a nitrate-based CIMS, also known as CI-API-TOF, for measuring sulfuric acid and more oxidized HOMs in the gas phase and an iodide-based CIMS equipped with FIGAERO focusing on detecting oxidation products of VOCs in the particle phase (see the Supplementary Materials). We estimated the HOM volatility from their thermal desorption temperature ($T_{\text{max}}$) together with the volatility parameterization developed by Tröstl and co-workers (14).
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