Molecular detection of *Acanthamoeba* spp., *Naegleria fowleri* and *Vermamoeba (Hartmannella) vermiformis* as vectors for *Legionella* spp. in untreated and solar pasteurized harvested rainwater

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**Abstract**

**Background:** *Legionella* spp. employ multiple strategies to adapt to stressful environments including the proliferation in protective biofilms and the ability to form associations with free-living amoeba (FLA). The aim of the current study was to identify *Legionella* spp., *Acanthamoeba* spp., *V. vermiformis* and *N. fowleri* that persist in a harvested rainwater and solar pasteurization treatment system.

**Methods:** Pasteurized (45 °C, 65 °C, 68 °C, 84 °C and 93 °C) and unpasteurized tank water samples were screened for *Legionella* spp. and the heterotrophic plate count was enumerated. Additionally, ethidium monoazide quantitative polymerase chain reaction (EMA-qPCR) was utilized for the quantification of viable *Legionella* spp., *Acanthamoeba* spp., *V. vermiformis* and *N. fowleri* in pasteurized (68 °C, 74 °C, 84 °C and 93 °C) and unpasteurized tank water samples, respectively.

**Results:** Of the 82 *Legionella* spp. isolated from unpasteurized tank water samples, *Legionella longbeachae* (35 %) was the most frequently isolated, followed by *Legionella norrlandica* (27 %) and *L. rowbothamii* (4 %). Additionally, a positive correlation was recorded between the heterotrophic plate count vs. the number of *Legionella* spp. detected (ρ = 0.710, P = 0.048) and the heterotrophic plate count vs. the number of *Legionella* spp. isolated (ρ = 0.779, P = 0.0028) from the tank water samples collected. Solar pasteurization was effective in reducing the gene copies of viable *V. vermiformis* (3-log) and *N. fowleri* (5-log) to below the lower limit of detection at temperatures of 68–93 °C and 74–93 °C, respectively. Conversely, while the gene copies of viable *Legionella* and *Acanthamoeba* were significantly reduced by 2-logs (P = 0.0024) and 1-log (P = 0.0015) overall, both organisms were still detected after pasteurization at 93 °C.

**Conclusions:** Results from this study indicate that *Acanthamoeba* spp. primarily acts as the vector and aids in the survival of *Legionella* spp. in the solar pasteurized rainwater as both organisms were detected and were viable at high temperatures (68–93 °C).

**Keywords:** Rainwater harvesting, Solar pasteurization, *Legionella*, *Acanthamoeba*, *Vermamoeba*, *Naegleria*
Background
The demand on fresh water supplies is intensifying as a result of an increase in the world’s population and urbanization, coupled with the negative effects of climate change [1–3]. Domestic rainwater harvesting systems can be utilized to augment existing surface- and groundwater supplies and in many countries it is utilized as a primary potable water source as well as for domestic and irrigation purposes. Several studies have however, highlighted that rainwater may become contaminated, especially during the harvesting process when debris, animal excreta, dust and leaves, which have accumulated on the roof catchment surface, are washed into the rainwater storage tank [4–6]. It is thus recommended that harvested rainwater is disinfected as numerous pathogens, including those that are opportunistic in nature, have previously been detected in this water source and are of a human health concern [7–12].

Dobrowsky et al. [13], indicated that a closed-coupled solar pasteurization system operating at temperatures greater than 72 °C can be utilized to treat harvested rainwater as the level of heterotrophic bacteria, Escherichia coli (E. coli) and total coliforms were reduced to below the detection limit and were within the respective drinking water guidelines [14–16]. However, PCR assays confirmed the presence of Yersinia spp., Legionella spp., and Pseudomonas spp., at temperatures greater than 72 °C, with Legionella spp. persisting at temperatures greater than 90 °C. A follow-up study performed by Reyneke et al. [17], indicated that Legionella spp. may be entering a viable but non-culturable (VBN) state as intact Legionella cells were detected at temperatures of up to 95 °C using ethidium monoazide (EMA) quantitative PCR (EMA-qPCR).

Legionella spp. exhibit a number of mechanisms enabling them to withstand environmental stresses such as heat treatment. These include associations with at least 20 protozoan hosts including Acanthamoeba spp., Naegleria spp., Vermamoeba (Hartmannella) vermiformis and Vahlkampfia spp. and two species of ciliated protozoa, including Tetrahymena spp. and Cyclidium spp. [18]. Their association with free-living amoeba (FLA) is especially effective as the amoeba host provides nutrients including, amino acids for the proliferation of Legionella spp. and a protective environment when Legionella spp. are enclosed in the cysts of the amoeba species [19, 20].

Of the genera belonging to the FLA, Acanthamoeba spp., Naegleria fowleri and V. vermiformis are the most frequently isolated from water samples [21–23], including samples from hot water systems [19, 24, 25]. Moreover, Acanthamoeba spp. and N. fowleri are associated with human and animal infections, including amoebic keratitis and severe brain pathologies [26–29]. The life-cycles of these FLAs are then divided into two stages. First, in the form of a vegetative trophozoite, the organism is able to feed and replicate. Secondly, a cyst is formed under unfavourable environmental conditions and this allows the organism to withstand nutrient starvation, heat, cold, desiccation and biocidal treatments [29–31]. Although there is limited data regarding FLA resistance to various disinfection procedures, they are a potential risk to public health not only because of the transmission of the protozoa themselves, but because they harbour a range of microbial pathogens including Legionella spp., Listeria monocytogenes, Pseudomonas aeruginosa and Mycobacterium spp., amongst other species [29, 32, 33].

Generally during the process of phagocytosis, the amoeba will engulf avirulent bacterial cells and form a phagosome. The phagosome then fuses with the lysosome, containing lysozymes, which degrade the bacterial cells [34]. Although there are differences in host-cell trafficking processes amongst L. pneumophila, L. micdadei and L. longbeachae, virulent Legionella spp. have the ability to halt the phagosome-lyosomal degradation pathway of the amoeba. This implies that the phagosome (containing the Legionella) does not undergo sequential maturation and therefore does not ultimately fuse with degradative lysosomes [35, 36]. The Legionella instead form a Legionella containing vacuole by recruiting secretory vesicles from the endoplasmic reticulum exit sites and mitochondria of the amoeba to the plasma membrane of the Legionella containing vacuole [36–38]. Owing to the proteins of the Type IVB defect in organelle trafficking/intracellular multiplication (Dot/Icm) secretion system, that aid in the establishment and the preservation of the Legionella containing vacuole, Legionella are then able to proliferate in this protective rough endoplasmic reticulum-like compartment [38, 39]. Once nutrients within the Legionella containing vacuole become limiting, Legionella will kill the amoeba and escape, where after they either establish a new replicative niche within a new host or continue to survive as planktonic cells and/or within biofilms as sessile cells [39, 40]. It has then been suggested that the growth of Legionella within amoeba hosts in the environment, is required to select or maintain virulent strains of Legionella able to cause Legionnaires’ disease [18, 39].

Numerous Legionella spp. (e.g. L. pneumophila and L. longbeachae, amongst others) have been known to cause the acute potentially fatal form of pneumonia as part of a multisystem disease known as Legionnaires’ disease (also referred to as Legionellosis or Legion Fever) [41] or a milder form of pulmonary infection known as Pontiac fever, which is a flu like illness [42]. Since the outbreak of Legionnaires’ disease has previously been linked to roof-harvested rainwater systems [8, 43] and hot water distribution systems [44], the aim of the current study was to isolate and identify the primary Legionella spp. contaminating a harvested rainwater and a solar pasteurization
(SOPAS) system (used for the treatment of roof-harvested rainwater) and to identify possible vectors including *Acanthamoeba* spp., *V. vermiformis* and *N. fowleri*, enabling their resistance and persistence. The viability of the FLAs as well as *Legionella* spp. at temperatures greater than 68 °C was also determined using EMA-qPCR. The enumeration of heterotrophic bacteria was included as previous studies have suggested that a correlation may exist between the heterotrophic plate count and *Legionella*, as *Legionella* multiply in biofilms as a survival strategy in the environment [45–47]. Additionally, the heterotrophic plate count was utilized to monitor the change in the number of viable heterotrophic bacteria before and after pasteurization [48]. Unpasteurized and pasteurized rainwater samples treated at temperatures of 45 °C, 65 °C, 68 °C, 74 °C, 84 °C and 93 °C, were collected for the enumeration of the heterotrophic plate count and the isolation of *Legionella* spp. However, as previous research has indicated that the indicator counts are reduced to below the detection limit at temperatures greater than 70 °C [13, 17], the samples collected at temperatures of 68 °C, 74 °C, 84 °C and 93 °C only were utilized for the EMA-qPCR experiments.

**Methods**

**Sample site and collection**

The Apollo ™ solar pasteurization system previously described by Dobrowsky et al. [13] was utilized for the treatment of harvested rainwater stored in a polyethylene rainwater harvesting (RWH) tank (2000 l). The RWH tank and pasteurization system were installed at the Welgevallen experimental farm (33°56′ S, 18°52′ 6.08″ E), Stellenbosch University, Western Cape, South Africa, during July 2013. Samples (5 l) pasteurized at 45 °C, 65 °C, 68 °C, 74 °C, 84 °C and 93 °C were collected from the Apollo ™ solar pasteurization system during September and October 2015, with six corresponding unpasteurized tank water samples (5 l) collected from the connecting RWH tank. The pH and temperature of the tank water samples were recorded at the sampling site, using a handheld pH55 pH/temperature meter (Martini Instruments, North Carolina, USA) and an alcohol thermometer, respectively.

Ambient temperature for the Stellenbosch area during 2015 were obtained from the South African Weather Services (Pretoria, South Africa), while global horizontal irradiance (GHI; W/m²) data were obtained from Stellenbosch Weather Services, Engineering Faculty, Stellenbosch University (http://weather.sun.ac.za/).

**Enumeration of the heterotrophic plate count**

For the enumeration of the heterotrophic plate count, a serial dilution (1:10) was prepared for each unpasteurized and pasteurized tank water sample, respectively, and 100 μl of each undiluted and diluted (10⁻¹–10⁻⁵) sample was spread plated onto Reasoner’s 2A agar (R2A agar; Difco Laboratories, Detroit, Michigan, USA), with the plates incubated at 37 °C for up to 4 days in accordance with Standard Methods 9215 C, American Public Health Association [49].

**Isolation of Legionella spp. from pasteurized and unpasteurized tank water samples**

*Legionella* spp. were recovered from the pasteurized (45 °C, 65 °C, 68 °C, 74 °C, 84 °C and 93 °C) and unpasteurized tank water samples according to the procedure outlined by the Centers for Disease Control and Prevention [50]. Briefly, 500 ml of each sample collected from the domestic rainwater harvesting tank and from the pasteurization system, respectively, was filtered through a sterile GN-6 Metricel® S-Pack Membrane Disc Filter ( Pall Life Sciences, Michigan, USA) with a pore size of 0.45 μm and a diameter of 47 mm. The filtration flow rate was approximately ≥ 65 ml/min/cm² at 0.7 bar (70 kPa). The filters were then aseptically removed from the filtration system and were placed into sterile 50 ml centrifuge tubes containing sterile water (5 ml). If more than one filter was required, additional filters were added to the same sample tube. The centrifuge tubes were then vortexed to detach the cells from the filters. As *Legionella* detection and isolation may be hampered by the growth of non-*Legionella* background flora [51], the selective detection for the unpasteurized tank water samples was increased by pre-incubating the cell suspension at 50 °C for 30 min before cultivation [50]. Thereafter, 100 μl of the cell suspension was spread plated onto buffered charcoal yeast extract (BCYE) agar containing ACES [N-(2-acetamido)-2-aminoethanesulfonic acid] buffer/potassium hydroxide (1.0 g/l), ferric pyrophosphate (0.025 g/l), alpha-ketoglutarate (0.10 g/l) and L-cysteine HCL (0.04 g/l) (Oxoid, Hampshire, England), BCYE agar supplemented with glycine, vancomycin, polymyxin B and cycloheximide (GVPC) and charcoal yeast extract (CYE) agar base without supplements (Oxoid, Hampshire, England).

All plates were then incubated at 35 °C for approximately 10 days. Colonies exhibiting a convex and round with entire morphology that appeared on the BCYE and GVBC media and not on the CYE medium were selected for further analysis. To confirm the presence of presumptive *Legionella* spp., colonies were streaked onto BCYE agar and Nutrient Agar (NA; Merck, Gauteng, South Africa) as a preliminary identification strategy. Colonies that grew only on BCYE agar and not NA were presumptively classified as *Legionella* spp., and were utilized for further analysis.

**Total genomic DNA extractions from presumptive Legionella isolates**

Total genomic DNA (gDNA) extractions were performed for presumptive *Legionella* spp. isolated from the tank
water samples present on the BCYE and GVBC media and not on the CYE media. Before the extraction of DNA, presumptive *Legionella* colonies were inoculated into buffered yeast extract (BYE) broth supplemented with ACES buffer/potassium hydroxide (1.0 g/l), ferric pyrophosphate (0.025 g/l), alpha-ketoglutarate (0.10 g/l) and L-cysteine HCL (0.04 g/l) according to Edelstein & Edelstein [52]. Presumptive *Legionella* cultures were then incubated at ± 35 °C for 4 days. Total gDNA was extracted from the cultures using the boiling method previously described by Ndlovu et al. [53].

Ethidium monoazide (EMA) treatment and total gDNA extractions from pasteurized and unpasteurized tank water samples

For the detection and quantification of *Legionella* spp., *Acanthamoeba* spp., *N. fowleri* and *V. vermiformis* total gDNA was extracted from pasteurized (68 °C, 74 °C, 84 °C, 93 °C) and the corresponding unpasteurized tank water samples. For this, 1 l of each sample was first subjected to flocculation as previously described by Dobrowsky et al. [54]. Briefly, 2 ml of CaCl₂ (1 M) and 2 ml of Na₂HPO₄ (1 M) were added to each sample before the samples were stirred (room temperature) for 5 min. The samples were then concentrated by filtration as outlined above. For the detachment of cells from the filters, the filters were transferred to 4 ml citrate buffer (0.3 M, pH 3.5) and were vortexed. The filters were subsequently removed and the 4 ml suspension was centrifuged at 16 000× g for 10 min. After the removal of the supernatant the pellet was re-suspended in 1 ml sterile MilliQ water before EMA treatment. Ethidium monoazide (2.5 µg/ml) was added to the 1 ml of concentrated sample according to Delgado-Viscogliosi et al. [55] and Chang et al. [56]. The suspension was then vortexed vigorously and placed on ice for 10 min in the dark. To cross link the EMA to the naked DNA, the samples were kept horizontal on ice and were exposed to a 500 W halogen light for 15 min at a distance of 20 cm. Following centrifugation (16 000× g, 5 min), the supernatant was removed and the pellet was washed with 1 ml NaCl (0.85 %). The sample was centrifuged (16 000× g, 5 min) and the pellet was re-suspended in 700 µl lysis buffer. The extraction of total gDNA was then performed using the Soil Microbe DNA MiniPrep™ Kit (Zymo Research, Irvine, USA) according to manufacturer’s instructions.

Conventional PCR assays for the identification of *Legionella* isolates

For the identification of the presumptive *Legionella* iso-lates, DNA was extracted from each isolate as outlined above. The primer set LEG 225/LEG 858 was then utilized to amplify 634 bp of the 16S rRNA sequence as previously described by Miyamoto et al. [57] (Table 1). The PCR mix consisted of 10 µl of 5x Green GoTaq® Flexi Buffer (1x; Promega, Madison, USA), 4 µl MgCl₂ (2.0 mM; Promega), 0.5 µl of each dNTP (0.1 mM; Thermo Fischer Scientific, Waltham, USA), 2 µl of each PCR primer (LEG 225 and LEG 858; 0.4 µM), 0.3 µl of GoTaq® Flexi DNA Polymerase (1.5U; Promega) and 2 µl of template DNA. All conventional PCR mixtures consisted of a final volume of 50 µl. The PCR cycling parameters were as follows: initial denaturation at 95 °C (1.5 min) followed by 30 cycles of denaturation at 94 °C (10 s), annealing at 64 °C (1 min) and elongation at 74 °C (1 min). A final extension was included at 72 °C (10 min).

Quantification of viable *Legionella* spp., *Acanthamoeba* spp., *N. fowleri* and *V. vermiformis* in pasteurized and unpasteurized tank water samples

For the quantification of viable *Legionella* spp., *Acanthamoeba* spp., *N. fowleri* and *V. vermiformis* in pasteurized

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### Table 1 Primers and amplification conditions utilized in the current study for the identification and quantification of *Legionella* spp., *Acanthamoeba* spp., *Naegleria fowleri* and *Vermamoeba* (Hartmannella) *vermiformis* in pasteurized and unpasteurized tank water samples

| Organism                        | Primer name | Primer sequence (5′–3′) | Gene  | Amplification conditions                        | Reference          |
|---------------------------------|-------------|-------------------------|-------|------------------------------------------------|--------------------|
| *Legionella* spp. (Identification) | LEG 225     | AAGATTAGCCTGGCTCCGAT    | 16S rRNA (634) | 95 °C (1.5 min) followed by 30 cycles of 94 °C (10 s), 64 °C (1 min) and 74 °C (1 min). Final extension: 72 °C (10 min). | Miyamoto et al. [57] |
|                                  | LEG 858     | GTCACGTACTGCGTTGCT      | 95 °C (1 min) followed by 45 cycles of 95 °C (15 s), 60 °C (15 s) and 72 °C (11 s) | Herpers et al. [58] |
| *Legionella* spp.                | Leg F       | CTAATCGTGCTAGTCTTCGAC   | 225-55 rRNA (259) | 95 °C (1 min) followed by 45 cycles of 95 °C (15 s), 60 °C (15 s) and 72 °C (11 s) | Qvarnstrom et al. [59] |
|                                  | Leg R       | CAATCGGACGTCTTGAG       | 185 rDNA (180)  | 95 °C (1 min) followed by 45 cycles of 95 °C (15 s), 60 °C (1 min) and 72 °C (40 s) | Qvarnstrom et al. [59] |
| *Acanthamoeba* spp.              | AcantF900   | CCCAGATCGTTACCGTCAA     | 185 rDNA (153)  | 95 °C (1 min) followed by 45 cycles of 95 °C (15 s), 60 °C (1 min) and 72 °C (40 s) | Qvarnstrom et al. [59] |
|                                  | AcantR1100  | TAAATTTAATGCCCCCAAATATCC |             | 95 °C (1 min) followed by 45 cycles of 95 °C (15 s), 60 °C (1 min) and 72 °C (40 s) | Qvarnstrom et al. [59] |
| *Naegleria fowleri*               | NaegF192    | GTGGTGAACCTAGCTGTTAATGCA | 185 rDNA (153)  | 95 °C (1 min) followed by 45 cycles of 95 °C (15 s), 60 °C (1 min) and 72 °C (40 s) | Qvarnstrom et al. [59] |
|                                  | NaegR344    | CACTAGAAAAAGCTACCTGAA   |             | 95 °C (3 min) followed by 45 cycles of 95 °C (20 s), 58 °C (30 s) and 72 °C (40 s) | Qvarnstrom et al. [59] |
| *Vermamoeba* (Hartmannella)     | Hv1227F     | TTACGAGGGCCGACCTG      | 185 rDNA (502)  | 95 °C (3 min) followed by 45 cycles of 95 °C (20 s), 58 °C (30 s) and 72 °C (40 s) | Kuiper et al. [60]  |
| *vermiformis*                     | Hv1728R     | GACCATCCCGAGTCTCG       |             | 95 °C (3 min) followed by 45 cycles of 95 °C (20 s), 58 °C (30 s) and 72 °C (40 s) | Kuiper et al. [60]  |
(68 °C, 74 °C, 84 °C, 93 °C) and unpasteurized tank water samples, quantitative PCR (qPCR) was performed using a LightCycler ® 96 (Roche, Gauteng, South Africa) following EMA treatment. For all qPCR assays, to a final reaction volume of 20 μl, using the FastStart Essential DNA Green Master Mix (Roche Applied Science, Mannheim, Germany), the following were added: 10 μl FastStart Essential DNA Green Master Mix (2×), 5 μl template DNA, and 0.4 μl of each primer (0.2 μM). For the quantification of Legionella spp. in pasteurized and unpasteurized tank water samples, the primers and qPCR parameters according to Herpers et al. [58] were utilized (Table 1). To generate a standard curve for the quantification of Legionella spp., the purified conventional PCR product obtained by amplifying the 256 bp product from L. pneumophila ATCC 33152 was utilized.

For the quantification of Acanthamoeba spp. in pasteurized and unpasteurized tank water samples, the primers and qPCR parameters as previously described by Qvarnstrom et al. [59] were utilized (Table 1). To generate the standard curve for the quantification of Acanthamoeba spp., the 180 bp PCR product amplified from gDNA of A. mauritaniensis ATCC 50677 was cloned into the pGEM T-easy vector system (Promega Corp.) according to the manufacturer's instructions. Once the plasmid had been sequenced, the plasmid containing the correct insert was used to generate the standard curve. Additionally, for the quantification of N. fowleri in pasteurized and unpasteurized tank water samples, the primers and qPCR parameters as outlined by Qvarnstrom et al. [59] were utilized (Table 1). To generate the standard curve for the quantification of N. fowleri a purified 153 bp PCR product obtained by screening a 1 l tank water sample from a domestic rainwater harvesting tank located at Stellenbosch University (GPS coordinates: 33°55′1.1″S, 18°51′56.7″E) using the Hv1227F/Hv1728R primer set was cloned into the pGEM T-easy vector system (Promega Corp.) according to the manufacturer's instructions. Once the plasmid had been sequenced, the plasmid containing the correct insert was used to generate the standard curve for the quantification of N. fowleri.

For the quantification of V. vermiformis in pasteurized and unpasteurized tank water samples, the primers and qPCR parameters according to Kuiper et al. [60] were utilized (Table 1). To generate a standard curve for the quantification of V. vermiformis, the purified conventional PCR product (502 bp) obtained by screening a 1 l tank water sample from a domestic rainwater harvesting tank located at Stellenbosch University (GPS coordinates: 33°55′5.1″S, 18°51′56.7″E) using the Hv1227F/Hv1728R primer set was utilized.

The concentration of the purified PCR products (Legionella spp. and V. vermiformis) and plasmid DNA (Acanthamoeba spp. and N. fowleri) were quantified using the NanoDrop™ ND-1000 (Nanodrop Technologies Inc., Wilmington, Delaware, USA) in triplicate at CAF. Serial 10-fold dilutions (10⁹ to 10¹) of the sequenced conventional PCR products and plasmid DNA were prepared in order to generate the standard curves for each respective organism. A concentration of 1.00 × 10⁹ gene copies/μl was prepared for the dilution with the highest copy number and a concentration of 1.00 × 10¹ gene copies/μl was prepared for the dilution with the lowest copy number. Standard curves generated by plotting quantitative cycle (Cq) values vs the log concentrations of standard DNA as previously described by Chen and Chang [61], were then used to determine the number of gene copies of each of the organisms. Melt curve analysis was included for all SYBR green real-time PCR assays in order to verify specificity of the primer set by ramping the temperature from 65 to 97 °C at a rate of 0.2 °C/s with continuous fluorescent signal acquisition at 5 readings/°C.

Sequencing of PCR amplicons
The PCR amplicons of each presumptive Legionella isolate, the PCR products used as positive controls to generate the standard curves for each qPCR assay and representative products of each of the qPCR assays of each organism were then purified using the DNA Clean & Concentrator™-5 Kit (Zymo Research) and were sent for sequencing at the CAF, Stellenbosch University. Chromatograms of each sequence were examined as outlined in Dobrowsky et al. [12] and all sequences were submitted as a query to BLAST for a sequence similarity search against the NCBI databases (https://blast.ncbi.nlm.nih.gov/Blast.cgi).

Inter- and intra-assay reproducibility and the lower limit of detection
To establish the inter-assay reproducibility of the qPCR assays optimized for each respective organism, the coefficient of variation (CV) was determined using the concentrations of nine dilutions (10⁹ to 10¹) of conventional PCR products and plasmid DNA that were quantified in duplicate during three separate qPCR experiments. In addition, the CV for intra- assay repeatability was calculated using the concentrations of nine dilutions (10⁹ to 10¹) for each qPCR assay [62–65]. In order to eliminate any PCR inhibitors, samples resulting in end-point fluorescence (EPF) values of less than 3.15 were diluted (10×) and the qPCR experiment was repeated for these samples. The minimum number of gene copies (highest dilution) that could be measured accurately within an assay was considered the lower limit of detection for each organism [63].

Statistical analysis
The data obtained from the microbial and physical analysis of the tank water samples collected, were assessed using the Statistical software package, Statistica™ version
13.0 (Statsoft Inc.). Before the data analysis, the Shapiro-Wilk test was used to test the normality of data sets. The gene copies of Legionella spp., Acanthamoeba spp., V. vermiformis and N. fowleri obtained for the pasteurized and unpasteurized tank water samples were assessed for nonparametric differences using the Mann-Whitney U test. Thus, the temperature of the tank water samples, before and after pasteurization was used as a single, ordinal variable. Spearman’s rank (p) correlation tests were performed to establish correlations between different microbiological (the heterotrophic plate count, number of Legionella isolates, the gene copies of Legionella spp., Acanthamoeba spp., V. vermiformis and N. fowleri obtained before and after pasteurization) and physical parameters (pH and temperature of tank water samples) as previously described by Wang et al. [66]. Significance was set at a P-value of ≤0.05 for all statistical analyses performed.

Results

Physical parameters of pasteurized and unpasteurized tank water samples collected during the sampling period (September–October 2015)

Pasteurized (45 °C, 65 °C, 68 °C, 74 °C, 84 °C and 93 °C) tank water samples were collected from the Apollo™ solar pasteurization system with corresponding unpasteurized tank water samples collected from the RWH tank during September and October 2015. The average daily ambient temperature ranged from 15.7 °C (September 2015) to 18.5 °C (October 2015). Additionally, as the Apollo™ solar pasteurization system relies on radiation from the sun to heat the tank water, the average total GHI was recorded at 8288.9 W/m² during September 2015 and 11574.6 W/m² during October 2015. The temperature of the water samples collected from the RWH tank ranged from the lowest temperature of 18 °C (15.09.2015) to the highest recorded temperature of 31 °C (27.10.2015). An average pH of 8.0 (range: 7.9–8.1) was recorded for unpasteurized tank water samples which then increased to pH 8.3 (range: 8.2–8.5) after pasteurization (Table 2).

The heterotrophic plate count and culturing of Legionella spp

For all unpasteurized tank water samples (n = 6), the heterotrophic plate count numbers ranged from 2.7 × 10⁵ CFU/ml to 1.5 × 10⁶ CFU/ml and were above the Department of Water Affairs and Forestry (DWAF) [14] guideline of 100 CFU/ml (Table 2). Additionally, the heterotrophic plate count were above the DWAF [14] guidelines following pasteurization at 45 °C (1.5 × 10⁵ CFU/ml) and 65 °C (4.7 × 10⁵ CFU/ml), respectively (results not shown). However, after the pasteurization treatment for the temperatures ranging from 68 to 93 °C, heterotrophic plate counts were reduced to below the detection limit (< 1 CFU/ml) and were within the DWAF guidelines (Table 2).

Conventional PCR for the identification of Legionella isolates

Culture based methods were then utilized to isolate Legionella spp. from all pasteurized (n = 6) and unpasteurized (n = 6) tank water samples. While no Legionella spp. were isolated from pasteurized tank water samples (45 °C to 93 °C), Legionella spp. were isolated from all the unpasteurized tank water samples utilizing culturing methods (Table 2). A total of 82 Legionella isolates were obtained overall from all the unpasteurized samples and all the resulting DNA sequences of the Legionella isolates displayed similarities to sequences of Legionella spp. recorded on NCBI. Legionella longbeachae (n = 29; GenBank accession no: FN650140.1, JN606078.1, NR_102800.1) was the species most frequently isolated from the unpasteurized tank water samples (results not

Table 2  Microbiological parameters and physical parameters determined for pasteurized and unpasteurized harvested rainwater samples

| Date        | Unpasteurized sample temp. (°C) | pH  | Heterotrophic plate count (CFU/ml) | No. of Legionella isolates obtained | Gene copies/ml | LS | AS | VS | NF |
|-------------|---------------------------------|-----|-----------------------------------|------------------------------------|----------------|----|----|----|----|
|             |                                 |     |                                   |                                    |                |    |    |    |    |
| 22.10.2015  | 24                              | 8.4 | 1.5 × 10⁶                          | 9                                  | 6.5 × 10⁴      | 9.8 × 10⁴ | 5.7 × 10⁵ | 1.6 × 10⁵ |     |
|             | 68a                             | 8.4 | BDLb                              | BDLb                               | 3.2 × 10³      | 8.4 × 10³ | 9.4 × 10³ | LLODc |     |
| 22.10.2015  | 25                              | 8.1 | 1.0 × 10⁶                          | 2                                  | 4.5 × 10⁴      | 8.3 × 10⁴ | 3.9 × 10⁴ | 6.4 × 10⁴ |     |
|             | 74a                             | 8.5 | BDLb                              | BDLb                               | 9.2 × 10³      | 5.2 × 10³ | LLODc | LLODc |     |
| 19.10.2015  | 21                              | 8.3 | 6.8 × 10⁴                          | 1                                  | 5.7 × 10⁵      | 1.3 × 10⁵ | 1.4 × 10⁵ | 9.2 × 10⁴ |     |
|             | 84a                             | 8.2 | BDLb                              | BDLb                               | 2.3 × 10³      | 1.7 × 10⁴ | LLODc | LLODc |     |
| 27.10.2015  | 31                              | 8.4 | 2.7 × 10⁵                          | 3                                  | 8.2 × 10⁴      | 6.5 × 10⁴ | 3.2 × 10⁵ | 1.0 × 10⁵ |     |
|             | 93a                             | 8.2 | BDLb                              | BDLb                               | 1.1 × 10³      | 1.4 × 10⁵ | LLODc | LLODc |     |

aPasteurized rainwater sample
bBDL below detection limit
cLLOD lower limit of detection: Vermamoeba vermiformis (< 5–8 gene copies/µl)
dLLOD lower limit of detection: Naegleria fowleri (< 12–17 gene copies/µl)
shown), followed by *Legionella norrlandica* (*n* = 22) and *Legionella rowbothamii* (*n* = 3) (Fig. 1; accession numbers included), the remaining 28 isolates were undetermined *Legionella* species (results not shown) and BLAST analysis indicated the presence of uncultured *Legionella* spp. (GenBank accession no: HQ111985.1, HQ111937.1, GU185995.1) and *Legionella* spp. (GenBank accession no: JN380993.1, JN380988.1).

### Quantification of viable *Legionella* spp. present in pasteurized and unpasteurized tank water samples

As the heterotrophic plate count was reduced to below the detection limit following pasteurization at 68 °C to 93 °C, viable *Legionella* spp. were quantified in samples pasteurized at 68 °C, 74 °C, 84 °C, 93 °C and the four corresponding unpasteurized tank water samples (Table 2). For all unpasteurized tank water samples, the concentration of viable *Legionella* ranged from $4.5 \times 10^4$ gene copies/ml (25 °C) to $8.2 \times 10^6$ gene copies/ml (31 °C) (Table 2). After pasteurization treatment, the highest concentration of *Legionella* was detected at 74 °C ($9.2 \times 10^3$ gene copies/ml) which then decreased to $1.1 \times 10^3$ gene copies/ml following pasteurization at 93 °C as indicated in Table 2. The number of *Legionella* gene copies then decreased by 1-log (87.2 %) following pasteurization at 68 °C and 74 °C, respectively. In contrast, a 3-log reduction (> 99.9 %) in *Legionella* gene copies was observed following pasteurization at 84 °C, while a 4-log (> 99.9 %) reduction in *Legionella* gene copies was observed following pasteurization at 93 °C. Overall, the number of viable *Legionella* gene copies decreased significantly ($Z = 3.034; P = 0.0024$) by an average of 2-logs (93.6 %) following pasteurization at 68–93 °C. Representative qPCR products were sequenced and *Legionella anisa* (GenBank accession no: JN001853.1, Z24682.1) and *Legionella monrovica* (GenBank accession no: Z24729.1) were detected in representative pasteurized and unpasteurized tank water samples following BLAST analysis, respectively.

### Quantification PCR efficiency, reproducibility and lower limit of detection

A linear range of quantification from $10^9$ to $10^1$ gene copies per μl of DNA extracts was observed for all standard curves produced for the quantification of *Legionella* spp., *Acanthamoeba* spp., *V. vermiformis* and *N. fowleri*, respectively. As indicated in Table 3, the qPCR assays had amplification efficiencies that ranged from 1.86 to 1.94 (*Legionella* spp.), 1.92 to 1.95 (*Acanthamoeba* spp.), 1.85...
to 1.89 (V. vermiformis) and 1.90 to 2.04 (N. fowleri). The optimum amplification efficiency is measured at 2.00 and corresponds to a doubling of copy number for every PCR cycle [67]. The correlation coefficient ($r^2$) ranged from 0.99 to 1.00 for all qPCR assays performed for Legionella spp., Acanthamoeba spp. and N. fowleri and 0.98 to 1.00 for all qPCR assays performed V. vermiformis, respectively (Table 3). The qPCR lower limit of detection was recorded at 8–12 gene copies/μL for Legionella spp., 2–6 gene copies/μL for Acanthamoeba spp., 5–8 gene copies/μL for V. vermiformis and 12–17 gene copies/μL for N. fowleri (Table 3). The qPCR assays demonstrated good reproducibility as the mean inter-and intra-assay coefficient of variation (CV) values and standard deviations (SD) were less than 1 and 5% for all qPCR assays, respectively (Table 3).

### Quantification of FLAs present in pasteurized and unpasteurized tank water samples

Following EMA treatment and DNA extractions, qPCR was performed for the quantification of viable Acanthamoeba spp., V. vermiformis and N. fowleri in samples pasteurized at 68 °C, 74 °C, 84 °C, 93 °C and the corresponding unpasteurized tank water samples, respectively (Table 2). As indicated in Table 2, gene copies of Acanthamoeba ranged from 6.5 x 10^4 gene copies/ml (31 °C) to 1.3 x 10^5 gene copies/ml (25 °C) for all unpasteurized tank water samples. After pasteurization treatment, the gene copies of Acanthamoeba spp. decreased and ranged from 5.2 x 10^3 gene copies/ml (74 °C) to 1.7 x 10^4 gene copies/ml (84 °C) and 1.4 x 10^5 gene copies/ml (93 °C) (Table 2). Overall, the number of Acanthamoeba gene copies decreased significantly ($Z = -3.183; P = 0.0015$) by 1-log (87.3%) following pasteurization at 68–93 °C. Acanthamoeba genotype T4 (GenBank accession no: KT892923.1), genotype T15 (GenBank accession no: KT892848.1) and Acanthamoeba lenticulata (GenBank accession no: KX018047.1) were detected in representative pasteurized and unpasteurized tank water samples, respectively.

For all unpasteurized tank water samples, the gene copies of viable V. vermiformis ranged from 3.9 x 10^4 gene copies/ml (25 °C) to 5.7 x 10^6 gene copies/ml (24 °C) (Table 2). Following pasteurization, the gene copies of V. vermiformis decreased and ranged from 9.4 x 10^3 gene copies/mL (68 °C) to below the lower limit of detection (< 5–8 gene copies/μL) for the remainder of the pasteurized samples (74 °C to 93 °C). A 3-log reduction (99.9%) in V. vermiformis gene copies was observed following pasteurization at 68 °C while a 5-log reduction (> 99.9%) in gene copies of viable V. vermiformis was observed following pasteurization at 74 °C to 93 °C, respectively. Overall, the number of gene copies of V. vermiformis (GenBank accession no: KT185625.1) decreased significantly ($Z = 3.067; P = 0.0021$) by 5-log (> 99.9%) following pasteurization at 68–93 °C.

For all unpasteurized tank water samples, gene copies of viable N. fowleri ranged from 6.4 x 10^4 gene copies/ml (25 °C) to 1.0 x 10^6 gene copies/ml (31 °C) (Table 2). Following pasteurization, the gene copies of N. fowleri decreased to below the lower limit of detection (< 12–17 gene copies/μL) for all tank water samples pasteurized at 68 °C, 74 °C, 84 °C and 93 °C, respectively (Table 2). A 5-log reduction (> 99.9%) in N. fowleri gene copies was observed following pasteurization at 68 °C to 84 °C, respectively, while a 6-log reduction (> 99.9%) in gene copies of N. fowleri was observed following pasteurization at 93 °C. Overall, the number of gene copies of viable N. fowleri (GenBank accession no: IQ271702.1, IQ271704.1) decreased significantly ($Z = 3.308; P = 0.001$) by 5.2-log (> 99.9%) following pasteurization at 68–93 °C.

### Associations of microbiological parameters and abiotic factors

As indicated in Table 4, Spearman’s ($\rho$) correlations were noted between parameters measured throughout the study. For example, positive correlations were observed between the heterotrophic plate count and the number of Legionella isolates obtained ($\rho = 0.779$, $P = 0.0028$) and the gene copies of Legionella spp. ($\rho = 0.710$, $P = 0.048$), Acanthamoeba spp. ($\rho = 0.862$, $P = 0.006$), V. vermiformis ($\rho = 0.858$, $P = 0.006$) and N. fowleri ($\rho = 0.810$, $P = 0.015$), respectively. Additionally, significant positive correlations were observed between the number of Legionella isolates obtained vs the gene copies of Legionella spp. ($\rho = 0.812$, $P = 0.0014$), Acanthamoeba spp. ($\rho = 0.761$, $P = 0.028$),
V. vermiformis (ρ = 0.936, P = 0.001) and N. fowleri (ρ = 0.946, P = 0.0004), respectively. The number of Legionella spp. gene copies/ml were also positively correlated to the gene copies of the amoeba detected including V. vermiformis (ρ = 0.854, P = 0.001), N. fowleri (ρ = 0.913, P = 0.002) and Acanthamoeba spp. (ρ = 0.643, P = 0.085). Moderate to high correlations were then detected between the number of Acanthamoeba spp. gene copies vs the number of V. vermiformis (ρ = 0.756, P = 0.03) and N. fowleri gene copies (ρ = 0.845, P = 0.028), respectively. A high correlation was also established between the gene copies of V. vermiformis and of N. fowleri (ρ = 0.936, P = 0.001).

**Discussion**

The heterotrophic plate count for all unpasteurized and tank water samples pasteurized at 45 °C and 65 °C were above the DWAF guideline for drinking water, with the heterotrophic plate count reduced to below the detection limit (< 1 CFU/ml) following pasteurization at 68 °C to 93 °C. The heterotrophic plate count represents only the culturable portion of the general bacterial community that are present in a water source and results indicated that pasteurization treatment at temperatures above 68 °C are effective as heterotrophic plate count counts in the treated samples corresponded to drinking water guidelines [14]. Previously, Sommer et al. [68] indicated that faecal coliforms were inactivated in river water at temperatures above 70 °C utilizing solar pasteurization. Moreover, Dobrowsky et al. [13] indicated that a closed-coupled solar pasteurization system operating at temperatures above 72 °C reduced the level of the heterotrophic plate count, E. coli and total coliforms to below the detection limit in harvested rainwater. In the current study a positive correlation between the heterotrophic plate count and the number of culturable Legionella present and the gene copies of viable Legionella spp. was also established. These results are in agreement with a study conducted by Serrano-Suárez et al. [69], where Legionella spp. were isolated when the corresponding heterotrophic plate count concentrations were above 1 × 10^5 CFU/100 ml, indicating that the frequency at which Legionella spp. are isolated may depend on the presence of culturable heterotrophic bacteria. Additionally, positive correlations were established between the heterotrophic plate count and the number of gene copies of viable Acanthamoeba spp., V. vermiformis and N. fowleri. This is expected as these FLAs are heterotrophs and known grazers of bacteria and the heterotrophic plate count represents the general bacterial microbiota [66, 69, 70]. Furthermore, while results indicated that the pH of the unpasteurized and pasteurized tank water samples did not significantly influence the microbiological quality, an increase in temperature of the pasteurized harvested rainwater significantly reduced the heterotrophic plate count, the number of Legionella isolates obtained, and the gene copies of viable Legionella spp., Acanthamoeba spp., V. vermiformis and N. fowleri detected, respectively.

Legionella spp. were also isolated using culture based techniques and during the current study, the majority of the Legionella isolates were obtained from unpasteurized tank water samples at temperatures of 18 °C (52 Legionella isolates) and 19 °C (15 Legionella isolates), respectively. Molecular analysis of the Legionella isolates obtained then indicated that L. longbeachae was the dominant Legionella spp. isolated from the unpasteurized tank water samples. While L. longbeachae is generally isolated from soil, including potting soil [71], this microorganism has previously been isolated from water samples collected from hospital reticulation systems and cooling towers and is able to proliferate in Acanthamoeba polyphaga [72–74]. In addition, while L. pneumophila serotype 1 is responsible for most of the human reported infections, 17 additional species have also been associated with disease and these include L. longbeachae, L. micdadei, L. anisa and L. bozemani [75]. BLAST analysis also revealed that L. norrlandica was of the dominant Legionella spp. isolated from unpasteurized tank water samples. This Legionella spp. harbours the majority of the L. pneumophila virulence factors and has only

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**Table 4** Spearman rank order correlation coefficients (ρ) of parameters investigated in this study

| Parameter                        | Temperature (°C) | pH       | HPC/ml   | No. of isolates (Culture) | Legionella spp. | Acanthamoeba spp. | Vermamoeba vermiformis |
|----------------------------------|-----------------|----------|----------|--------------------------|-----------------|-------------------|------------------------|
| Temperature (°C)                 | –               | –        | –        | –                        | –               | –                 | –                      |
| pH                              | 0.847**         | 0.367    | –        | –                        | –               | –                 | –                      |
| HPC/ml                          | 0.885**         | 0.779**  | 0.710*   | 0.812*                   | 0.756*          | –                 | –                      |
| No. of isolates (Culture)        | 0.833**         | 0.756*   | 0.862**  | 0.761*                   | 0.643           | –                 | –                      |
| Legionella spp.                  | –               | 0.118    | 0.810*   | 0.946**                  | 0.913**         | –                 | 0.936**                |
| Acanthamoeba spp.                | –               | 0.118    | 0.810*   | 0.946**                  | 0.913**         | –                 | 0.936**                |
| Naegleria fowleri                | –               | 0.118    | 0.810*   | 0.946**                  | 0.913**         | –                 | 0.936**                |

*P < 0.05, **P < 0.01
recently been described, where Rizzardi et al. [76] isolated a novel Legionella genus from the biopurification systems of wood processing plants. Moreover, the study group revealed that L. norrlandica could establish a replicative vacuole in A. castellanii. Three isolates were also identified as Legionella rowbothamii. Adeleke et al. [77] reported on the characterization of a novel Legionella spp., namely L. rowbothamii, and to date no studies have reported on the isolation of L. rowbothamii from environmental samples. Moreover, no studies have indicated whether L. rowbothamii proliferates in protozoa. As L. rowbothamii was isolated during the current study, future research should thus elucidate whether L. rowbothamii is able to colonize and proliferate in amoeba species.

Although Legionella spp. were isolated from the unpasteurized tank water samples, no Legionella spp. were detected using the culture based methods in the pasteurized tank water samples (45 °C to 93 °C). Numerous studies have indicated that temperatures below 50 °C are not sufficient to eradicate Legionella spp. from water distribution systems [13, 17, 46, 69] and it is therefore unexpected that Legionella spp. were not isolated from particularly the 45 °C pasteurized tank water sample. However, in agreement with previous studies that have focused on the thermal inactivation of Legionella spp., the culturability of Legionella from the pasteurized tank water samples may have been affected by the heat treatment and by the nutrient shock of going from a nutrient poor environment such as rainwater, onto the nutrient rich environment provided by the media, as this induces Legionella cells to enter a viable but non-cultivable (VBNC) state [78, 79].

EMA-qPCR assays were then performed for all unpasteurized and tank water samples pasteurized at 68 °C, 74 °C, 84 °C and 93 °C to determine whether Legionella and the FLA’s were viable. Although viable V. vermiformis was detected in all the unpasteurized tank water samples, results indicated that solar pasteurization at 74–93 °C was effective in reducing the gene copies of V. vermiformis to below the lower limit of detection (< 5–8 gene copies/μl). Additionally, N. fowleri were not detected in any of the pasteurized tank water samples and results of the current study thus indicate that the thermal treatment of tank water at 68–93 °C is sufficient for the removal of N. fowleri as the gene copies of viable N. fowleri were reduced to below the lower limit of detection (< 12–17 gene copies/μl) for all pasteurized tank water samples. Although discrepancies, such as the presence of multicellular communities, may arise when analysing environmental samples, Fouque et al. [31] indicated that V. vermiformis cysts were completely inactivated at 70 °C, which is in agreement with the current study, where V. vermiformis was more sensitive to heat treatment than Acanthamoeba spp. Additionally, previous studies have indicated that Naegleria spp. are considered more sensitive to heat treatments compared to thermotolerant Acanthamoeba and V. vermiformis [24, 80, 81]. Results obtained in the current study do however, indicate that rainwater harvesting tanks are vulnerable to N. fowleri and V. vermiformis colonization and amoeba, including N. fowleri and V. vermiformis, should be included in the surveillance of pathogens in drinking water distribution systems [82].

Viable Legionella spp. and Acanthamoeba spp. were then detected in all the pasteurized and unpasteurized tank water samples. It has been well established that Legionellae are facultative intracellular parasites of amoeba including, Acanthamoeba spp., Naegleria spp. and V. vermiformis [83]. Although significant positive correlations were observed between the number of Legionella gene copies and the gene copies of V. vermiformis and N. fowleri, it is hypothesized that Legionella spp. may primarily be associating with Acanthamoeba spp. during thermal treatment as viable Legionella spp. and Acanthamoeba spp. persisted in all pasteurized (68 °C to 93 °C) and unpasteurized tank water samples. This is not surprising, as previous studies have detected Legionella at high pasteurization temperatures (> 90 °C) in solar pasteurization systems using molecular based techniques, including EMA-qPCR [13, 17]. Although this has not been demonstrated for all Legionella spp., L. pneumophila has been known to survive and proliferate on the debris of dead microbial cells such as heat-killed Pseudomonas putida, E. coli, Bacillus subtilis, Lactobacillus plantarum, A. castellanii and Saccharomyces boulardii [84]. Moreover, Acanthamoeba spp. are able to graze on heat killed bacteria including E. coli and Klebsiella spp. [70]. In agreement with Thomas et al. [81], it is thus hypothesized that the solar pasteurization system may be indirectly providing favourable conditions for Legionella and Acanthamoeba spp. and these organisms may thus be surviving on the dissolved organic constituents, available through the decay of the microorganisms at high pasteurization temperatures. However, while the presence of dissolved organic constituents may allow for the survival of Legionella spp., it is hypothesized that Acanthamoeba cysts may be harbouring Legionella and allow the Legionella spp. to proliferate and grow in harvested rainwater and during the treatment process. This is in agreement with Storey et al. [85] who indicated that Acanthamoebae cysts remained viable after heat treatment at 80 °C for 10 min. It is further hypothesized that during the DNA extraction process the cysts may be lysed and Legionella are released and detected using molecular methods including EMA-qPCR. Additionally, the qPCR assays utilized in the current study indicated high reproducibility as the mean inter-assay CV values and SD were less than 5 and 1 %, respectively, which were comparable to CV and SD values obtained by Ahmed et al. [65].
The current study therefore highlights the value of EMA-qPCR for the detection of viable *Legionella* and their protozoan hosts as opposed to culture based techniques that may yield false negative results [82].

**Conclusions**

Although incidences of Legionnaires’ disease are well documented for regions including Europe, the USA, New Zealand and Australia, limited information is available on the environmental distribution of *Legionella* spp. as well as incidences of Legionnaires’ disease in developing countries such as South Africa [83, 86–88]. The surveillance of *Legionella* in water distribution systems is thus vital as *Legionella* have been described as “new or emerging pathogens in drinking water” [89]. The current study demonstrated that culture-based methods for the detection of *Legionella* are less sensitive and with the use of EMA-qPCR, viable *Legionella* spp., *Acanthamoeba* spp., *V. verminiformis* and *N. fowleri* were detected in untreated tank water samples, while viable *Legionella* spp. (93 °C), *Acanthamoeba* spp. (93 °C) and *V. verminiformis* (68 °C) were detected after pasteurization.

Additionally, insight into the presence and persistence of *Legionella* spp., and amoeba including *Acanthamoeba*, *V. verminiformis* and *N. fowleri* in a representative rainwater harvesting tank and a solar pasteurization treatment system was provided. The occurrence of these pathogens in harvested rainwater is of particular concern as they are frequently detected in water distribution systems and residential plumbing [45]. For example, *N. fowleri* is the causative agent of the disease, primary amoebic meningoencephalitis (PAM) and although PAM infections are rare, the mortality rate is extremely high [90].

The presence of viable *Legionella* spp. and *Acanthamoeba* spp. highlights the need for further investigation as solar pasteurization may be insufficient for the long-term control of pathogenic *Acanthamoeba* and *Acanthamoebae*-bound *Legionella* in harvested rainwater.

**Abbreviations**

ACES: N-(2-acetamido)-2-aminoethanesulfonic acid; BYE: Buffered yeast extract; CAF: Central analytical facility; CV: Coefficient of variation; CYE: Charcoal yeast extract; Dot/Icm: Defect in organelle trafficking/intracellular multiplication; DWAF: Department of Water Affairs and Forestry; EMA: Ethidium monoazide; EMA-qPCR: Ethidium monoazide quantitative polymerase chain reaction; EPF: End-point fluorescence; FLA: Free-living amoeba; gDNA: Genomic DNA; GHI: Global horizontal irradiance; GVPC: Glycine, vancomycin, polymyxin B and cycloheximide; NA: Nutrient agar; NCBI: National centre for biotechnology information; pPCR: Quantitative PCR; R2A: Reasoner 2A agar; RWH: Rainwater harvesting; SD: Standard deviation; VBNC: Viable but non-culturable

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**Availability of data and materials**

The datasets during and/or analysed during the current study is available from the corresponding author on reasonable request.

**Authors’ contributions**

WK, SK and PD conceived and designed the experiments. PD performed the experiments and analyzed the data. WK and TE acquired funding for the study; WK and SK contributed reagents/materials/analysis tools. WK and PD wrote the paper. All authors edited the drafts of the manuscript and approved the final version of the manuscript.

**Competing interests**

The authors declare that they have no competing interests.

**Consent for publication**

Not applicable.

**Ethics approval and consent to participate**

Not applicable.

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**References**

1. Murad AA, Al Nuaimi H, Al HM. Comprehensive assessment of water resources in the United Arab Emirates (UAE). Water Resour Manag. 2007;21(9):1449–60.

2. Wheldr E, Verhoeven R. An alternative solution of the water shortage problem in Libya. Water Resour Manag. 2007;21(6):561–82. doi:10.1007/s11269-006-9067-6.

3. O’Hara JK, Georgakakos KP. Quantify the urban water supply impacts of climate change. Water Resour Manag. 2008;22(10):1477–97. doi:10.1007/s11269-008-9238-8.

4. Ahmed W, Hodgers L, Siddhu JPS, Toze S. Fecal indicators and zoonotic pathogens in household drinking water taps fed from rainwater tanks in Southeast Queensland, Australia. Appl Environ Microbiol. 2012;78(1):219–26.

5. Lee JY, Bak G, Han M. Quality of roof-harvested rainwater—comparison of different roofing materials. Environ Pollut. 2012;162:422–9.

6. De Kwaadsteniet M, Dobrowsky PH, Van Deventer A, Khan W, Cloete TE. Domestic rainwater harvesting: microbiological and chemical water quality and point-of-use treatment systems. Water Air Soil Poll. 2013;224(7):1–19.

7. Albrechtsen H. Microbiological investigations of rainwater and graywater collected for toilet flushing. Water Sci Technol. 2002;46(6–7):311–6.

8. Simmons G, Jury S, Thornley C, Harte D, Mohiuddin J, Taylor M. A Legionnaires’ disease outbreak: A water blaster and roof-collected rainwater systems. Water Res. 2008;42(6):1449–58.

9. Ahmed W, Huggens F, Goonetilleke A, Gardner T. Real-time PCR detection of pathogenic microorganisms in roof-harvested rainwater in Southeast Queensland, Australia. Appl Environ Microbiol. 2008;74(7):5490–6.

10. Ahmed W, Gardner T, Toze S. Microbiological quality of roof-harvested rainwater and health risks: a review. J Environ Qual. 2011;40(1):13–21.

11. Dobrowsky PH, Mannel D, De Kwaadsteniet M, Prozesky H, Khan W, Cloete TE. Quality assessment and primary uses of harvested rainwater in Kleinmond, South Africa. Water SA. 2014;40(3):401–6.
12. Dobrowsky PH, De Kwaadsteniet M, CioeTE, Khan W. Distribution of indigenous bacterial pathogens and potential pathogens associated with roof-harvested rainwater. Appl Environ Microbiol. 2014;80(7):2307–16.

13. Dobrowsky PH, Carstens M, De Villiers J, CioeTE, Khan W. Efficiency of a closed-coupled solar pasteurization system in treating roof harvested rainwater. Sci Total Environ. 2015;536:206–14.

14. Department of Water Affairs and Forestry (DWAF). South African water quality guidelines 2nd Edition Volume 1: Domestic water use. Pretoria: CSIR Environmental Services; 1996.

15. NHMRC, NRMMC. Australian drinking water guidelines (ADWG). Paper 6, national water quality management strategy. Canberra: National Health and Medical Research Council, National Resource Management Ministerial Council Commonwealth of Australia; 2011.

16. South African Bureau of Standards (SABS). South African national standards (SANS) 241: Drinking Water Quality Management Guide for Water Services Authorities. Annexure A. ISBN 0-626-17752-9. 2005.

17. Reyneke B, Dobrowsky PH, Ndlovu T, Khan S, Khan W. EMA-qPCR to monitor the efficiency of a closed-coupled solar pasteurization system in reducing Legione\ell\oe contamination of roof-harvested rainwater. Sci Total Environ. 2016;566:70–2.

18. Lau HY, Ashbolt NJ. The role of biofilms and protozoa in Legionella pathogenesis: implications for drinking water. J Appl microbiol. 2009;107(2):368–78.

19. Thomas V, Herrera-Rimann K, Blanc DS, Greub G. Biodiversity of amoebae and amoeba-resistant bacteria in a hospital water network. Appl Environ Microbiol. 2006;72(8):4238–8.

20. Magnét A, Peralta R, Gomes T, Izquierdo F, Fernandez-Vadillo C, Galvan A, et al. Vectorial role of Acanthamoeba in Legionella propagation in water for human use. Sci Total Environ. 2015;505:889–95.

21. Liu H, Ha YR, Lee ST, Hong YC, Yong HH, Chung DI. Genetic diversity of Acanthamoeba isolated from ocean sediments. Korean J Parasitol. 2006;44(2):117–25.

22. Thomas V, Loret JF, Jousset M, Greub G. Biodiversity of amoebae and amoeba-resistant bacteria in a drinking water treatment plant. Environ Microbiol. 2008;10(10):2728–45.

23. Rodríguez-Zaragoza S. Ecology of free-living amoebae. Crit Rev Microbiol. 1994;20(3):225–41.

24. Rohr U, Weber S, Michel R, Selenka F, Wilhelm M. Comparison of free-living amoebae in hot water systems of hospitals with isolates from moist sanitary areas by identifying genera and determining temperature tolerance. Appl Environ Microbiol. 1998;64(5):1822–4.

25. Buse HY, Lu J, Stueweing IT, Ashbolt NJ. Eukaryotic diversity in premise drinking water using 18S rDNA sequencing: implications for health risks. Environ Sci Pollut Res Int. 2013;20(9):6351–63.

26. Nagington JP, Watson TJ, Playfair J, McGill Jones BR, Steele ADM. Amoebic infection of the eye. Lancet. 1974;304(7896):1537–40.

27. John DT. Opportunistically pathogenic free-living amoebae. In: Kreier JP, Baker JR, editors. Parasitic protozoa. San Diego, California: Academic Press, Inc.; 1999. p. 143–246.

28. Marciano-Cabral F, Cabral G. Acanthamoeba spp. as agents of disease in humans. Clin Microbiol Rev. 2003;16(2):273–307.

29. Coulon C, Collignon A, McDonnell G, Thomas V. Resistance of Acanthamoeba cysts to disinfection treatments used in health care settings. J Clin Microbiol. 2010;48(8):2697–9.

30. Neff RJ, Ray SA, Benton WF, Wilborn M. Methods in cell physiology. In: Prescott DM, editor. Induction of synchronous encystation (differentiation) in Acanthamoeba spp. New York: Academic; 1964. p. 55–83.

31. Fouque E, Héchard Y, Hartemann P, Humeau P, Roy CR, Modulation of Rab GTPase function by a protein phosphoconelate transfe. Nature. 2001;477(7362):103–6.

32. Brüggemann H, Cazorla C, Buchhieser C. Adaptation of Legionella pneumophila to the Host environment: role of protein secretion, effectors and eukaryotic-like proteins. Curr Opin Microbiol. 2006;9(1):86–94.

33. Asare R, Santic M, Gabin I, Doric M, Sutlles J, Graham J, et al. Genetic susceptibility and caspase activation in mouse and human macrophages are distinct for Legionella longbeachae and L. pneumophila. Infect Immun. 2007;75(4):1933–45.

34. Horwitz MA. The Legionnaires’ disease bacterium (Legionella pneumophila) inhibits phagosome-lysosome fusion in human monocytes. J Exp Med. 1983;158(6):2108–20.

35. Cazalet C, Gomez-Valero L, Runiok C, Lomma M, Denvirs-Ravault D, Newton HJ, et al. Analysis of the Legionella longbeachae genome and transcriptome uncovers unique strategies to cause Legionnaires’ disease. PLoS Genet. 2010;6(2):e1000851.

36. Allombert J, Fuche F, Michaud C, Douillet P. Molecular mimicry and original biochemical strategies for the biogenesis of a Legionella pneumophila replicative niche in phagocytic cells. Microb Infect. 2013;15(1):491–8.

37. Molofsky AB, Swanston MS. Differentiate to thrive: lessons from the Legionella pneumophila life cycle. Mol Microbiol. 2004;53(1):29–40.

38. Fraser DW, Tsai TR, Orenstein W, Parkin WE, Beecham H, Sharrar RG, et al. Legionnaires’ disease: description of an epidemic of pneumonia. N Engl J Med. 1977;297(22):1189–90.

39. Glick TH, Gregy MB, Berman B, Mallison G, Rhodes WWJ, Kassanoff I. Pontiac fever. An epidemic of unknown etiology in a health department. I. Clinical and epidemiologic aspects. Am J Epidemiol. 1978;107(2):149–60.

40. Schlech WF, Gorman GW, Payne MC, Broome CV. Legionnaires’ disease in Caribbean: an outbreak associated with a resort hotel. Arch Intern Med. 1985;145(11):1076–9.

41. Dubreucq SF, Locas MC, Duerwieser A, Ruester C, Ismail J, Lefebvre B, et al. Sporadic Legionnaires’ disease: the role of domestic electric hot-water tanks. Epidemiol Infect. 2012;140(1):172–81.

42. Borella P, Guerrieri E, Marchesi I, Bondi M, Messi P. Water ecology of Legionella and protozoan: environmental and public health perspectives. Biotechnol Annu Rev. 2005;11:355–80.

43. Mouchouti V, Veloniski E, Tsakalof A, Kapoula C, Goutziana G, Vatopoulos A, et al. Risk factors for contamination of hotel water distribution systems by Legionella species. Appl Environ Microbiol. 2007;73(3):1489–92.

44. Duda S, Baron JL, Wagener MM, Vidic RD, Stout JE. Lack of correlation between Legionella colonization and microbial population quantification using heterotrophic plate count and adenosine triphosphate bioluminescence measurement. Environ Monit Assess. 2015;187(7):1–10.

45. Carter JT, Rice EW, Buchberger SG, Lee Y. Relationships between levels of heterotrophic bacteria and water quality parameters in a drinking water distribution system. Water Res. 2000;34(5):1495–502.

46. American Public Health Association (APHA). Standard methods for the examination of water and wastewater. Washington, DC: American Water Works Association and Water Pollution Control Federation; 1996.

47. Centers for Disease Control and Prevention (CDC). Procedures for the recovery of Legionella pneumophila from the environment. National Center for Infectious Diseases. Division of Bacterial and Mycotic Diseases, Respiratory Disease Laboratory Section. US: US Department of Health and Human Services; 2005.

48. Schets FM, Italiaander R, Van Den Berg HJUL, de Roda Husman AM. Rainwater harvesting: quality assessment and utilization in The Netherlands. J Water Health. 2010;8(2):224–35.

49. Edelstein PH, Edelstein MA. Comparison of three buffers used in the formulation of buffered charcoal yeast extract medium. J Clin Microbiol. 1993;31(12):3329–30.

50. Ndlovu T, Le Roux M, Khan W, Khan S. Co-detection of virulent Escherichia coli genes in surface water sources. PLoS One. 2015;10(2):e0116808.

51. Dobrowsky PH, Lombard M, Coete WJ, Saayman M, Coete TE, Carstens M, et al. Efficiency of microfiltration systems for the removal of bacterial and viral contaminants from surface and rainwater. Water Air Soil Pollut. 2015;226(3):1.

52. Delgado-Visconti P, Solignac L, Delattre J-M. Viability PCR a culture-independent method for rapid and selective quantification of viable Legionella pneumophila cells in environmental water samples. Appl Environ Microbiol. 2009;75(11):3502–12.

53. Chang CW, Lu LW, Kuo CL, Hung NT. Density of environmental Acanthamoeba and their responses to superheating disinfection. Parasitol Res. 2013;112(1):3687–96.

54. Miyamoto H, Yamato H, Arima K, Fuji J, Maruta K, Ito K, et al. Development of a new semi nested PCR method for detection of Legionella species and its application to surveillance of Legionelle in Hospital cooling tower water. Environ Microbiol. 1997;3(7):2489–94.
58. Herpers BL, de Jongh BM, van der Zwaluw K, van Hannen EJ. Real-time PCR assay targets the 23S-SS spacer for direct detection and differentiation of Legionella spp. and Legionella pneumophila. J Clin Microbiol. 2003;41(10):4815–6.

59. Qvamstrøm Y, Vivesvera GS, Srimat R, da Silva AJ. Multiplex real-time PCR assay for simultaneous detection of Acanthamoeba spp., Balanomastix mandrillas, and Naegleria fowleri. J Clin Microbiol. 2006;44(10):3589–95.

60. Kuper MW, Valster RM, Wulfsing BA, Boonstra H, Smidt H, Van Der Kooy D. Quantitative detection of the free-living amoeba Hartmannella vermiciformis in surface water by using real-time PCR. Appl Environ Microb. 2006;72(9):4750–6.

61. Chen NT, Chang CW. Rapid quantification of viable Legionellae in water and biofilm using ethidium monoazide coupled with real-time quantitative PCR. J Appl Microbiol. 2010;109(2):323–34.

62. Hellemans J, Mortier G, De Paepe A, Speleman F, Vandesompele J. qBase relative quantification software and framework for management and automated analysis of real-time quantitative PCR data. Genome Biol. 2007;8(2):R19.

63. Schmittgen TD, Livak KJ. Analyzing real-time PCR data by the comparative CT method. Nat Protoc. 2008;3(6):1101–8.

64. Bustin SA, Benes V, Garson JA, Hellemans J, Huggett J, Kubista M, et al. The MIQE guidelines: minimum information for publication of quantitative real-time PCR experiments. Clin Chem. 2009;55(4):601–22.

65. Ahmed W, Brandes H, Gyawali P, Sidhu JPS, Toze S. Opportunistic pathogens in roof-captured rainwater samples, determined using real-time PCR. Water Res. 2014;53:361–9.

66. Wang H, Edwards M, Falkingham JO, Pruden A. Molecular survey of the occurrence of Legionella spp., Mycobacterium spp., Pseudomonas aeruginosa and amoeba hosts in two chlorinated drinking water distribution systems. Appl Environ Microb. 2012;78(17):6285–94.

67. Ulrich MP, Norwood DA, Christensen DR, Ulrich RL. Using real-time PCR to specifically detect Burkholderia mallei. J Med Microbiol. 2006;55(5):551–9.

68. Sommer B, Mannino A, Solarte Y, Salas ML, Dieroff C, Valiente C, et al. SCODIS – an emerging water treatment process. J Water SRT – AQUA. 1997;46(3):127–37.

69. Serrano-Suárez A, Dellundé J, Salvadó H, Cervero-Aragó S, Méndez J, Canals O, et al. Microbial and physicochemical parameters associated with Legionella contamination in hot water recirculation systems. Environ Sci Pollut Res Int. 2013;20(8):5534–44. doi:10.1007/s11356-013-1557-5.

70. Weeks PH, Bodeler PL, Wijen JP, Vogels GD. Effects of grazing by the free-living soil amoebae Acanthamoeba castellanii, Acanthamoeba polyphaga, and Hartmannella vermiformis on various bacteria. Appl Environ Microb. 1993;59(7):2317–9.

71. Velonakis EN, Kiousi IM, Koutis C, Papadogiannakis E, Babatsikou F, Vatopoulos A. First isolation of Legionella pneumophila serogroup 1, in Greek potting soils: possible importance for public health. Clin Microbiol Infect. 2010;16(6):763–7.

72. Hartmannella vermiformis. Microbiological and public health implications of Naegleria fowleri in water distribution pipeline biofilms and drinking water samples. Environ Sci Technol. 2009;43(17):6691–6.

73. Puzon GV, Lancaster JA, Wylie JT, Plumb L. Rapid detection of Naegleria fowleri in water distribution pipeline biofilms and drinking water samples. Appl Environ Microb. 2006;72(6):4323–8.

74. Temmerman R, Vervaeren H, Noseda B, Boon N, Verstraete W. Necrotrophic growth of Legionella pneumophila. Appl Environ Microb. 2006;72(6):4323–8.

75. Storey MV, Wiencek-Krusnell J, Ashbolt NJ, Stenstrom TA. The efficacy of heat and chlorine treatment against tolerant Legionella pneumophila and Legionella. Scand J Infect Dis. 2004;36(9):656–62.

76. Barrie C, Venter S, Neil L. Identification methods for Legionella from environmental samples. Water Res. 2003;37(6):1362–70.

77. Deacon R. Prevention of Legionella strains in cooling towers and legionellosis cases in New Zealand. J Environ Health. 2013;75(6):682.

78. Graham FF, White PS, Harte DJG, Kingham SP. Changing epidemiological trends of legionellosis in New Zealand 1979–2009. Epidemiol Infect. 2012;140(8):1481–96. doi:10.1017/S0950268811000975.

79. Szewyck J, Zurowski R, Marcz W, Schleifer KH. Microbiological safety of drinking water. Annu Rev Microbiol. 2006;60(1):111–127.

80. Marciano-Cabral F, Cabral GA. The immune response to Naegleria fowleri amebae and pathogenesis of infection. FEMS Immunol Med Micr. 2007;51(2):243–59.