INTRODUCTION

In all countries, the amount of municipal waste is constantly increasing, contributing to a significant deterioration of the natural environment [Raharjo et al., 2017; Husár et al., 2019; Nagendra et al., 2019]. The awareness of how important the care for the world around us is reaches a growing group of people. The selective collection of municipal waste (paper, glass, plastics, metals) has become widely used in households and workplaces. First, waste generation should be prevented or reduced; if it is not possible, the waste is recovered. The use of other recovery processes is only implemented when it is not possible to prepare it for reuse or recycling for technological reasons or it is not justified for ecological or economic reasons. The wastes that cannot be recovered should be disposed of, for example by incineration. This process reduces the volume of municipal waste disposed in landfills and enables the recovery of energy. However, the incineration of municipal solid waste causes the formation of secondary products, i.e. secondary wastes in the form of ashes of various types (bottom or fly ashes). The management of this material in a balanced manner is now an important issue because it contains heavy metals [Wielgosiński et al., 2014].

The goal of this study was the estimation of the fractional composition of heavy metals usually present in municipal solid waste incineration (MSWI) bottom ash, such as copper, cadmium and lead [Joseph et al., 2018] in particle size fraction ≤2mm, which contains the largest amounts of metals [Chimenos et al., 1999]. The metal content in fractions better reflects the environmental threats related to heavy metals, as compared to the total content. It describes which part of the total metal content can be mobilized due to the changes of the environmental factors (e.g. pH, redox potential, etc.). Heavy metals, depending on
the content in individual forms are more or less mobile, i.e. bioavailable. In order to investigate the metal content in fractions, including cadmium, lead and copper in bottom ash, the sequential extraction method is used [Yao et al., 2011; Gonzales et al., 2019; Haberl and Schuster, 2019]. There are various procedures for the fractionation of metals, although one of the most economical ones is the BCR method (proposed by Standards, Measurement and Testing Programme, former Community Bureau of Reference).

MATERIALS AND METHODS

Sampling

The fresh bottom ash samples from MSWI plant in Białystok (North Eastern Poland) were collected for a period of seven weeks in the year 2016. During this period, 11650 tons of wastes were incinerated (average calorific value 8.95 MJ·kg⁻¹). The wastes were incinerated in the stoker-fired furnace (temperature on the grate – 700°C). The mixed municipal waste (from containers) and the waste after the sorting process are subjected to the thermal disposal.

Twenty samples were taken, approximately 1 kg each. The bottom ash was water quenched and separated magnetically to recover ferrous metals before sampling. The samples were sieved through a 2 mm sieve and ground in an agate mortar.

Analysis of physicochemical properties

Dry mass was evaluated by drying at 105°C to the constant mass. Pseudo-total Cd, Pb and Cu content was determined by means of atomic absorption spectrometry with flame atomisation (FAAS), after previous microwave digestion (0.5 g of dry ash + 10 cm³ HNO₃ + 3 cm³ HClO₄). The pH was determined with the potentiometric method (measured after 24 hours in suspension – 5 g of ash in 25 cm³ of deionized water).

Sequential extraction procedure

The BCR method with the usage of ultrasonic probe Sonics VCX 130 was used to evaluate fractional composition of cadmium, lead and copper in the dry MSWI bottom ash samples. The extraction included four stages [Łukowski, 2017]:

1. The acid soluble and exchangeable fraction (F1) – 1 g of ash in 100 cm³ centrifuge tube with 40 cm³ of 0.11 mol·dm⁻³ acetic acid was sonicated for 7 minutes (power – 20W) at a temperature of 22±5°C. Then, the mixture was centrifuged for 20 minutes at 3000 g. The extract was separated for analysis. The residue with 20 cm³ of deionized water was sonicated for 5 minutes (power – 20W) and centrifuged for 20 minutes at 3000 g. Water was discarded.

2. The reducible fraction, bound to Fe/Mn oxides (F2) – 40 cm³ of 0.5 mol·dm⁻³ hydroxylamine hydrochloride fresh solution, pH 1.5 was added to the residue from the first step, and sonicated for 7 minutes (power – 20W) at a temperature of 22±5 °C. Then, the mixture was centrifuged for 20 minutes at 3000g. The extract was separated for analysis. The residue was rinsed with deionized water, like in the first step.

3. The oxidizable fraction, bound to organic matter (F3) – 20 cm³ of 30% hydrogen peroxide was added to the residue from the second step and sonicated for 2 minutes (power – 20W) at the temperature of 22±5°C. Then, the volume of H₂O₂ was reduced to approx. 1 cm³ using water bath. Afterwards, 50 cm³ of 1 mol·dm⁻³ ammonium acetate was added to the moist residue and sonicated for 6 min. (power – 20W) at a temperature of 22±5 °C. Then, the mixture was centrifuged for 20 minutes at 3000 g. The extract was separated for the analysis. The residue was rinsed with deionized water, like in the previous steps.

The determination of cadmium, copper and lead content in fractions was carried out by FAAS. The percentage of particular fractions of cadmium, copper and lead in the pseudo-total content and the percentage of mobile pool (F1 – F3) were calculated. The content of Cd, Cu and Pb in the residual fraction (F4) was calculated from the difference between the pseudo-total contents and the contents in the F1-F3 fractions.

RESULTS AND DISCUSSION

Physicochemical properties of MSWI bottom ash

The average dry mass content was 77.5% and ranged from 74.0 to 84.1 (Table 1). The obtained results are generally in accordance with those given by Wiles [1996], who stated that bottom ash contains usually from 15% to 25% of water...
and the moisture is essential to control fugitive dust and to reduce the volume of ash.

One of the most important chemical properties of ash is pH. It is one of the main factors influencing the metal leaching. The MSWI bottom ash utilization is mainly dependent on the hazards of heavy metal contamination [Quek et al., 2016]. The average pH value in the studied MSWI bottom ash was 11.79. A similar pH for the municipal solid waste incineration bottom ash is given by Yao et al. [2011] and Yao et al. [2012], 11.2 and 10.85 respectively. Moreover, according to Dijkstra et al. [2006], the fresh bottom ash samples are mainly alkaline (pH = 10–12).

The least pseudo-total content among the studied metals was noted in the case of cadmium (1.48 mgˑkg$^{-1}$) and the highest for copper (1135.0 mgˑkg$^{-1}$ DM on average). As stressed by Dou et al. [2017], based on the data from around the world, the MSWI bottom ash contains approximately 100–10000 mgˑkg$^{-1}$ of Cu and Pb as well as 1–100 mgˑkg$^{-1}$ of Cd. Chimenos et al. [1999], emphasized that copper and lead accumulate mainly in the 1–2 mm particle size fraction of municipal waste. According to these authors, the highest amounts of copper and lead in this fraction are associated with small parts of copper wires and alloys used for soldering. The wide range of pseudo-total metal contents in the studied bottom ash is probably caused by the variable composition of municipal waste.

**Fractionation of metals in MSWI bottom ash**

The F1 fraction accumulated the largest amount of cadmium, 0.55 mgˑkg$^{-1}$, on average (Table 2). It was 37.5% of pseudo-total content (Figure 1) at the range from 12.4 to 62.4%. As stated by Kouassi et al. [2014], under neutral or slightly acidic conditions, the metals in this fraction can be leached very easily. It means that the Cd contained in fresh MSWI bottom ash poses a major environmental and human health problem, despite its lowest pseudo-total content among the studied metals. This also emerges from the well-known high toxicity of cadmium and the large scale of bottom ash production. Such a high content in exchangeable and acid soluble fraction must result from the properties of cadmium, which is considered as volatile element. According to Yao et al. [2013], most of Cd in fine fraction of ash originates from evaporation. The evaporated Cd is reacting with CO$_2$ in the furnace, thus creating carbonates. Additionally, the mobile pool of cadmium (72.8% at the range from 37.2 to 97.2% of pseudo-total content), was the highest in comparison with the lead and copper. That intensifies the threats connected with the Cd leaching under changing environmental conditions. The metals present in the F2 fraction are dependent on the redox potential changes and can be mobilized under reducible conditions or by low level of oxygen. The metals from the F3 fraction would be released by oxidizable conditions [Pöykiö et al., 2016].

| Metal | Fraction | Min – Max | Mean ± SD, n=20 |
|-------|----------|-----------|-----------------|
| Cd (mg kg$^{-1}$ DM) | F1 | 0.11 – 1.57 | 0.55±0.36 |
|      | F2 | 0.03 – 1.17 | 0.37±0.29 |
|      | F3 | 0.02 – 0.84 | 0.15±0.18 |
|      | F4 | 0.01 – 1.62 | 0.40±0.48 |
| Pb (mg kg$^{-1}$ DM) | F1 | 0.94 – 230.76 | 24.63±54.98 |
|      | F2 | 14.38 – 190.23 | 51.91±43.80 |
|      | F3 | 11.45 – 127.93 | 44.59±30.58 |
|      | F4 | 6.10 – 236.38 | 69.72±147.5 |
| Cu (mg kg$^{-1}$ DM) | F1 | 2.49 – 163.79 | 47.37±31.29 |
|      | F2 | 5.92 – 224.48 | 84.56±58.45 |
|      | F3 | 43.98 – 543.15 | 251.06±146.12 |
|      | F4 | 66.97 – 1984.64 | 752.03±704.04 |

**Table 1.** Physicochemical characteristics of MSWI bottom ash (n=20)

| Bottom ash samples | DM | pH | Pseudo-total content [mgˑkg$^{-1}$ DM] |
|--------------------|----|----|--------------------------------------|
|                    |    |    | Cd | Pb | Cu |
| min – max          | 74.0 – 84.1 | 10.96 – 12.66 | 0.47 – 4.35 | 42.4 – 743.1 | 311.5 – 2342.1 |
| mean ± SD          | 77.5±3.1 | 11.79±0.47 | 1.48±1.04 | 190.9±147.5 | 1135.0±704.2 |

**Table 2.** The content of cadmium, lead and copper in the fractions in MSWI bottom ash
The F4 Fraction gathered the lowest amount of cadmium (27.2%) as compared to the rest of metals. It is the result of high volatility, mentioned above. It was confirmed by the percentage of lead (36.5%) and copper (66.3%) in this fraction, since Pb is less volatile than Cd [Zhou et al., 2019] and Cu is less volatile than Pb [Ke et al., 2017]. The share of Cd in the F2 fraction (25.3% on average) was similar in comparison with Pb (27.2% on average). Fraction F3 gathered 10.0% of Cd. It was the lowest amount of metal in this fraction among studied metals.

The highest amount of lead, 69.72 mgˑkg\(^{-1}\) on average, was found in the F4 fraction. It was 36.5% of pseudo-total content at the range from 4.6 to 69.0%. Additionally, Yao et al. [2010] confirmed the largest Pb percentage in the residual fraction. On the basis of the research on the MSWI bottom ash originated from six incineration plants in the Zhejiang province (China), the authors noted more than 50% of Pb in this fraction. The lead in the residual fraction is bound to the calcio-alumino-silicate phases [Gonzales et al. 2019]. The percentage of Pb in the F2 and F3 fractions was similar, 27.2 and 23.4% respectively. The content of lead in the fraction bound to the Fe/Mn oxides was the highest among studied metals. The mobile pool of Pb gathered 63.5% of pseudo-total content and ranged from 24.6 to 95.4%. It is in good accordance with the results of Pöykiö et al. [2016] who stated about 69.4% of lead in the F1-F3 fractions.

Similarly, like in the case of Pb, the largest content of copper, 752.03 mgˑkg\(^{-1}\) on average, was noted in the F4 fraction. That comprised 66.3% of pseudo-total content and ranged from 10.2 to 89.2%. It was also the highest share among all the fractions of the investigated metals.

According to Yao et al. [2013], the Cu in the MSWI bottom ash is mostly present in the residual fraction. For the bottom ash fractions of <0.45 mm, 0.45–1 mm and 1–2 mm they noted 69.2%, 64.6% and 66.3% of copper, respectively. The residual fraction is chemically stable and biologically inactive. It mainly includes the metals embedded in the crystal lattice of primary and secondary minerals which are the part of the ash from the incineration of municipal waste. It consists mainly of silicate and aluminosilicate minerals. Under natural conditions, excluding the influence of unfavourable external factors, they can be considered permanently immobilized. In the F1 fraction, we stated the lowest share of copper (4.2% on average). It was also the lowest amount among all the fractions of studied metals. The F2 fraction gathered 7.4% of pseudo-total copper, the least as compared to lead and cadmium. Organic matter bound fraction (F3) comprised 22.1% of pseudo-total copper on average. Copper is known from strong affinity to organic matter. It forms complexes with organic ligands, e.g. fulvic acid [Yao et al., 2011]. This is confirmed by the low amount of Cu in the first two fractions, as stated above. The most of the organic matter from municipal solid waste is damaged through the incineration process [Yao et al., 2010]. According to Chimenos et al. [1999], the unburned organic matter in the 1–2 mm fraction of MSWI bottom ash comprises below 4% of weight. In the mobile pool of Cu, we found only 33.7% of pseudo-total content at the range from 10.8 to 89.8%. It was the lowest total content in the F1-F3 fractions among investigated heavy metals, suggesting the least risk of metal leaching.

CONCLUSIONS

1. The highest percentage of metal in acid soluble and exchangeable fraction (most mobile) was noted in the case of cadmium.
2. The average cadmium content in particular fractions can be arranged in the following order: F1 (37.5%) > F4 (27.2%) > F2 (25.3%) > F3 (10.0%); for lead: F4 (36.5%) > F2(27.2%) > F3(23.4%) > F1(12.9%) and for copper: F4(66.3%) > F3(22.1%) > F2(7.4%) > F1(4.2%).
3. Assuming the summary percentage of metal in mobile fractions (F1-F3) as the solubility criterion, we can state that cadmium was the most soluble and copper the least soluble.
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