Effect of Persulfate Activation by Electrogenerated $\text{H}_2\text{O}_2$ and Anodic Oxidation on the Color Removal of Dye Solutions at Pt and BDD Anodes

Yifan Yao $^1$, Kai Zhu $^{1,*}$, Yucan Liu $^2$, Qianjin Liu $^1$ and Lihua Huang $^{1,*}$

1. Shandong Provincial Key Laboratory of Water and Soil Conservation and Environmental Protection, College of Resources and Environment, Linyi University, Linyi 276000, China
2. School of Civil Engineering, Yantai University, Yantai 264050, China
* Correspondence: zhukai@lyu.edu.cn (K.Z.); huanglihua328@163.com (L.H.)

Abstract: In this study, tartrazine solutions were oxidized using innovative electrochemical advanced oxidation processes (EAOPs) that combined persulfate (PS) activation with electrogenerated $\text{H}_2\text{O}_2$, cathodic reduction and anodic oxidation at Pt and BDD anodes, and graphite cathode in an undivided stirred reactor. For the Pt anode, $\text{SO}_4^{2-}$ was generated from a reduction reaction at the cathode and a reaction between the PS and electrogenerated $\text{H}_2\text{O}_2$. For the BDD anode, $\text{SO}_4^{2-}$ was generated from a reduction reaction at the cathode, an oxidation reaction at the anode, and a reaction between PS and electrogenerated $\text{H}_2\text{O}_2$. Among these activation methods, the activation efficiency of PS by electro-generated $\text{H}_2\text{O}_2$ is much better than other methods. The effects of PS concentration up to 36 mM, applied current density between 6 to 15 mA cm$^{-2}$ and temperatures between 25 to 45 °C were investigated. For the electro-Fenton process with Pt anode (Pt-$\text{H}_2\text{O}_2$/PS process), the best result for oxidizing 250 mg L$^{-1}$ tartrazine solution was obtained with 37.5 mM Na$_2$SO$_4$ + 9.0 mM Na$_2$S$_2$O$_8$, applied current density at 12 mA cm$^{-2}$ and 45 °C, acquiring total color removal after 30 min reaction. For the electro-Fenton process with BDD anode (BDD-$\text{H}_2\text{O}_2$/PS process), the best result for oxidizing 250 mg L$^{-1}$ tartrazine solution was obtained with 25 mM Na$_2$SO$_4$ + 18 mM Na$_2$S$_2$O$_8$, applied current density at 12 mA cm$^{-2}$ and 45 °C, yielding 100% color removal after 30 min reaction. The main oxidizing agents are $\text{SO}_4^{2-}$ and $\text{OH}^-$ in the anodic oxidation process with PS and the electro-Fenton process with PS. It is concluded that the additions of PS tremendously improve the oxidation power of electro-Fenton processes with PS, especially the Pt-$\text{H}_2\text{O}_2$/PS process.

Keywords: anodic oxidation; electrogenerated $\text{H}_2\text{O}_2$; persulfate; hydroxyl radical; sulfate radical

1. Introduction

Textile manufacturing industries produce large amounts of wastewater containing high concentrations (up to 250 mg L$^{-1}$) of soluble azo dyes [1]. These high-concentration contaminations cannot be significantly removed by conventional wastewater treatment plants owing to their stable chemical structure’s resistance to biodegradation [2]. As a result, many types of azo dyes have been detected in reservoirs, rivers, and even drinking water [3]. Although the carcinogenic, mutagenic, and teratogenic effects of most azo dyes remain unknown, heightened concerns still exist concerning the potential adverse consequences on living organisms, including human beings [4].

Recently, electrochemical advanced oxidation processes (EAOPs) have attracted increasing interest as promising technologies for the degradation of azo dyes, providing several characteristic advantages, such as easy operation, simple equipment, and ambient temperature and pressure [5,6]. As with the most popular and convenient EAOPs, anodic oxidation (AO) is used to degrade contaminations by directly oxidation through electron transfer and/or indirectly oxidation through electrogenerated reactive species (hydroxyl...
radical) from simultaneous oxidation of water molecules at the anode surface \( A \), as described in Equation (1) [7,8]. In reality, indirect oxidation mechanisms involving water intermediates are the convincing paradigm for the AO process at anodes based on recently developed superior boron-doped diamond (BDD) [9]. However, at an anode based on noble metal (Pt), A-OH\(^-\) can be further oxidized to higher oxide (A=O). Hence, at Pt anode, degradation is mediated electrolysis with organics (R) oxidized by A=O, regenerating the reduced anode surface \( A \), as described in Equations (2) and (3) [10]. During the AO process, the degradation efficiencies of azo dyes rely largely on the properties of the anode

\[
\text{H}_2\text{O} \rightarrow \text{A-OH}^- + \text{H}^+ + e^- (1)
\]

\[
\text{A-OH}^- \rightarrow \text{A=O} + \text{H}^+ + e^- (2)
\]

\[
\text{A=O} + \text{R} \rightarrow \text{R=O} + \text{A} (3)
\]

\[
\text{O}_2 + 2\text{H}^+ + 2e^- \rightarrow \text{H}_2\text{O}_2 (4)
\]

\[
\text{A} + \text{H}_2\text{O}_2 \rightarrow \text{A-HO}_2^- + \text{H}^+ + e^- (5)
\]

In fact, SO\(_4^{2-}\) oxidation (\( E^0 = 2.44 \) V/SHE) is considered a burgeoning AOPs with strong oxidation power, slightly milder than OH\(^-\) oxidation (\( E^0 = 2.80 \) V/SHE) [15]. During the AO/PS and AO-H\(_2\)O\(_2\)/PS processes, SO\(_4^{2-}\) can be formed through several representative activation methods, including thermal activation via Equation (6) [16], reduction at cathode via Equation (7) [17], reaction with H\(_2\)O\(_2\) or organics (R) via Equation (8) and via Equation (9) [18,19].

\[
\text{S}_2\text{O}_8^{2-} + \text{heat} \rightarrow 2\text{SO}_4^- (6)
\]

\[
\text{S}_2\text{O}_8^{2-} + e^- \rightarrow \text{SO}_4^{2-} + \text{SO}_4^{2-} (7)
\]

\[
\text{S}_2\text{O}_8^{2-} + \text{H}_2\text{O}_2 \rightarrow 2\text{SO}_4^- + 2\text{OH}^- (8)
\]

\[
\text{S}_2\text{O}_8^{2-} + \text{R} \rightarrow 2\text{SO}_4^- + \text{R}^- (9)
\]

Several studies have investigated the effect of PS activation in EAOPs, with dyes [20], pharmaceutical and personal care products (PPCPs) [21], and herbicides [22] as targeted contaminations. Surprisingly, little has been done to investigate the effect of PS activation on the degradation efficiency of the AO-H\(_2\)O\(_2\)/PS process, especially comparing the AO-H\(_2\)O\(_2\)/PS process with different anode materials. During the AO/PS and AO-H\(_2\)O\(_2\)/PS oxidation processes, the generation and yield of OH\(^-\) largely depend on the nature of anode materials [23]. The generated OH\(^-\) at the electrode surface via Equation (1) can activate PS via Equation (10) when BDD is used as an anode, while SO\(_4^{2-}\) from Equation (7) can also be activated to SO\(_4^{2-}\) by anodic oxidation via Equation (11). The combination of SO\(_4^{2-}\) and OH\(^-\) results in a complex reaction environment [24,25]. On the one hand, SO\(_4^{2-}\) activated by H\(_2\)O\(_2\) and reduction reaction at the cathode can improve the oxidation power of the AO/PS and AO-H\(_2\)O\(_2\)/PS processes. On the other hand, SO\(_4^{2-}\) generated at the anode
surface will consume OH· in the system and occupy active anode sites, competing with contamination oxidation.

$$\text{S}_2\text{O}_8^{2-} + \text{A-OH} \rightarrow \text{A} + \text{SO}_4^{2-} + \text{HSO}_4^-$$ (10)

$$\text{SO}_4^{2-} \rightarrow \text{SO}_4^{2-} + e^-$$ (11)

This work aims to assess the effect of PS activation by electrogenerated H$_2$O$_2$ and anodic oxidation on the degradation efficiencies of the AO/PS and AO-H$_2$O$_2$/PS processes with two typical anodes, Pt and BDD, using tartrazine azo dye as targeted contamination. Comparative trials were made in the present study to clarify the role of electrogenerated A-OH·, OH·, and SO$_4^{2-}$ on the decolorization efficiency and mineralization of tartrazine solution. The effects of PS concentration, current density, and temperature on the decolorization performance were evaluated.

2. Materials and Methods

2.1. Materials

Tartrazine (>95%) was purchased from Macklin (China) and used without any further purification. Sodium sulfate and sodium persulfate used as supporting electrolytes were obtained from Sinopharm Chemical Reagent Beijing Co., Ltd. (Beijing, China). Sulfuric acid and sodium hydroxide used for adjustment of solution pH were purchased from Sinopharm Chemical Reagent Beijing Co., Ltd. Analytical standards methanol and tert-butanol were purchased from Aladdin (China). BDD and the platinum plate used as the anode material was supplied by NeoCoat (Switzerland) and General Research Institute for Nonferrous Metals (Beijing, China), respectively. Graphite felt (GF), supplied by Jiuhua Carbon (Xiangtan, China), was used as a cathode. Distilled water obtained from Watsons was used to prepare all solutions in this study.

2.2. Electrochemical Oxidation Experiments

The electrochemical oxidation of tartrazine was performed in an undivided plexiglass-made reactor with a 300 mL capacity. During the oxidation process, tartrazine solutions with different electrolytes were under vigorous magnetic stirring in a water bath kettle to keep mass transfer and temperature. The electrochemical oxidation experiments were conducted under galvanostatic mode with an ITECH IT6333A (USA) to supply power. The reactor was equipped with either a Pt plate (2.5 cm × 5 cm) or a BDD electrode (2.5 cm × 5 cm) as the anode; the immersed area of all anodes was 10 cm$^2$. During the AO/PS process, the cathode was a graphite cathode with an immersed area of 10 cm$^2$ (2.5 cm × 5 cm), which was substituted by larger graphite (immersed area of 36 cm$^2$, 6 cm × 6 cm) fed with air at a flow rate of 1 L min$^{-1}$ to carry out the AO-H$_2$O$_2$/PS trials. The distance between the active sides of the anode and cathode was kept at about 1 cm, and the initial pH value of the working solution was adjusted to 6.8 with NaOH or H$_2$SO$_4$. Samples, filtered through a 0.22 µm filter, were taken during the oxidation process at certain time intervals to analyze the color removal.

2.3. Analytic Methods

The absorbance of filtered samples with 20 times dilution was determined by a Metash UV-8000 Spectrophotometer (Shanghai, China) with the detector $\lambda = 429$ nm using a standard quartz cuvette (1 cm of the optic path). The percentage of color removal was calculated as follows:

$$\text{Color Removal} = \frac{A_0 - A}{A_0} \times 100$$ (12)

The electrogenerated H$_2$O$_2$ was determined by a standard spectrophotometric method detecting the absorbance of yellowish Ti (IV) complex using a Metash UV-8000 Spectrophotometer (Shanghai, China) with the detector $\lambda = 408$ nm.
3. Results and Discussion

3.1. Effect of PS Concentration on Color Removal

The effect of PS concentration was evaluated by adding to the dye solutions four supporting electrolytes, including 50 mM Na$_2$SO$_4$ (0% PS), 37.5 mM Na$_2$SO$_4$ + 9.0 mM Na$_2$S$_2$O$_8$ (25% PS), 25 mM Na$_2$SO$_4$ + 18 mM Na$_2$S$_2$O$_8$ (50% PS) and 36 mM Na$_2$S$_2$O$_8$ (100% PS). It is worth noting that all these four supporting electrolytes have approximately the same conductivity, between 6.9 and 7.5 mS cm$^{-1}$ [20]. Figure 1 shows the color removal with different supporting electrolytes during the AO/PS process (Figure 1a Pt/PS, Figure 1b BDD/PS) and the AO-H$_2$O$_2$/PS process (Figure 1c Pt-H$_2$O$_2$/PS, Figure 1d BDD-H$_2$O$_2$/PS) at a constant current density of 9 mA cm$^{-2}$. Table 1 shows the related degradation data during oxidation process.

![Figure 1a](image1.png)

![Figure 1b](image2.png)

**Figure 1.** Cont.
Figure 1. Effect of PS concentration on color removal during the AO/PS and AO-H\textsubscript{2}O\textsubscript{2}/PS processes, (a) Pt/PS, (b) BDD/PS, (c) Pt-H\textsubscript{2}O\textsubscript{2}/PS, and (d) BDD-H\textsubscript{2}O\textsubscript{2}/PS. Raw dye solution: 250 mg L\textsuperscript{-1} tartrazine, initial pH: 6.8, applied current density: 9 mA cm\textsuperscript{-2}, reaction temperature: 25 °C.

| Oxidation System | 0% PS | 25% PS | 50% PS | 100% PS |
|------------------|-------|--------|--------|---------|
| Pt/PS            | 71    | 94     | 63     | 64      |
| BDD/PS           | 90    | 83     | 95     | 72      |
| Pt-H\textsubscript{2}O\textsubscript{2}/PS | 36    | 95     | 91     | 81      |
| BDD-H\textsubscript{2}O\textsubscript{2}/PS | 69    | 96     | 93     | 74      |

* For Pt/PS and BDD/PS, data of 30 min was selected. For Pt-H\textsubscript{2}O\textsubscript{2}/PS and BDD-H\textsubscript{2}O\textsubscript{2}/PS, data of 20 min was selected.
The anode materials have a crucial influence on the oxidation efficiency of EAOPs [26]. To systematically research the effect of PS activation on the oxidation power of AO/PS and AO-H₂O₂/PS, two representative anodes, BDD and Pt, were used in this study. For the AO/PS process, it can be observed from Figure 1a,b that the color removals increased with the increase in reaction time. For reaction time of 30 min and PS concentration of 0%, 25%, 50%, and 100%, the color removals were obtained as 71%, 94%, 63%, and 64%, respectively for Pt anode, 90%, 83%, 95%, and 72%, respectively for BDD anode. It is evident that the appropriate addition of PS can increase the oxidation power of the AO/PS process, with the optimal concentration of PS being 25% for Pt and 50% for BDD. This phenomenon can be ascribed to the generated SO₄²⁻ via Equations (6), (7), and (10). Nevertheless, excessive dosing of PS leads to an obvious decrease in decolorization because of the competitive effect between PS and tartrazine on the anode surface. For the Pt anode, the decolorization is a mediated electrolysis with tartrazine oxidized by A=O [27]; the generation of SO₄²⁻ largely depends on reduction at the cathode via Equation (7). However, for the BDD anode, the oxidation of tartrazine relies on the electrogenerated A-OH· [28], and SO₄²⁻ generated through Equations (7) and (10) promotes the color removal. Due to the disparate generation pathway of SO₄²⁻ and differences in properties of anode materials, distance exists between the optimal concentration of the Pt/PS and BDD/PS processes.

For the AO-H₂O₂/PS process, at a reaction time of 20 min and PS concentration of 0%, 25%, 50%, and 100%, the color removals were obtained as 36%, 95%, 91%, and 81%, respectively for the Pt anode, 69%, 96%, 93%, and 74%, respectively for the BDD anode, as shown in Figure 1c,d. It can be observed that the additions of PS tremendously improve the oxidation power of the AO-H₂O₂/PS process. Furthermore, compared to the BDD-H₂O₂/PS process, the promotion of oxidation power by the addition of PS is more obvious in the Pt-H₂O₂/PS process. The more distinct promotion effects of PS in the AO-H₂O₂/PS process may cause by the activation reaction between PS and H₂O₂ as described in Equation (8), yielding greater productions of SO₄²⁻ and OH· in the bulk of the reaction solution. Surprisingly, excessive dosing of PS still results in a decrease in decolorization in the AO-H₂O₂/PS process. This phenomenon confirms the governing factor of active radicals generation via Equation (8) is the concentration of H₂O₂. In the AO-H₂O₂/PS process, the accumulation of H₂O₂ is a relatively slow course, especially with a carbonaceous cathode without any modification [29]. In this study, almost total decolorization can be obtained at a reaction time of 45 min. The accumulated concentration of H₂O₂ was just 56 mg L⁻¹ in 50 mM Na₂SO₄ solution without tartrazine at a reaction time of 45 min. For the Pt-H₂O₂/PS process, adding PS is an effective strategy to improve color removal since the generations of SO₄²⁻ and OH· rely on Equation (8) to introduce large numbers of free radicals into the system. However, for the BDD-H₂O₂/PS process, this improvement effect is not obvious compared with the Pt-H₂O₂/PS process. This trend may be associated with the competitive effect between PS and tartrazine on the anode surface, as described in Equation (10). Therefore, 37.5 mM Na₂SO₄ + 9.0 mM Na₂S₂O₈ (25% PS) and 25 mM Na₂SO₄ + 18 mM Na₂S₂O₈ (50% PS) were chosen as the optimal supporting electrolytes to investigate the influence of applied current density for the Pt and BDD anode, respectively.

3.2. Effect of Applied Current Density on Color Removal

Current density, the applied current carried by per unit area of the electrode, is the most vital parameter in EAOPs [30]. In the AO/PS and AO-H₂O₂/PS processes, current density regulates electron transfer and the production of OH· at the anode and H₂O₂ at the cathode [31]. Figure 2 shows the effect of applied current density ranging from 6 to 15 mA cm⁻² on the color removal of tartrazine solutions in the AO/PS process and the AO-H₂O₂/PS process. Table 2 shows the related degradation data during oxidation process.
Figure 2. Cont.
For the Pt/PS process, at a reaction time of 20 min, PS concentration of 25% and applied current densities of 6, 9, 12, and 15 mA cm\(^{-2}\), the color removals were obtained as 51%, 76%, 97%, and 96%, respectively, as shown in Figure 2a. Furthermore, for the BDD/PS process, at a reaction time of 20 min, PS concentration of 50%, and applied current densities of 6, 9, 12, and 15 mA cm\(^{-2}\), the color removals were obtained as 68%, 83%, and 100%, respectively, as shown in Figure 2b. Generally, a gradual increase in decolorization can be noticed with the increase of the current density from 6 to 12 mA cm\(^{-2}\). For the Pt/PS process, this phenomenon may be caused by the progressive generation of SO\(_4^{2-}\) via Equation (7) at the cathode and A=O at the anode. For the BDD/PS process, the promotion effect may be ascribed to the growing yield of A-OH\(_2\) via Equation (1) at the anode and SO\(_4^{2-}\) via Equations (7) and (10). However, the color removal of tartrazine solution remained stable with the current density increased from 12 to 15 mA cm\(^{-2}\) in the Pt/PS process and the BDD/PS process. The abovementioned tendency can be attributed to the parasitic reactions at the anode and the cathode at high applied current density [32]. With the increase of current density, the hydrogen evolution reaction at the cathode and oxygen evolution reaction at the anode surface are enhanced, competing with the production of OH\(^-\) and SO\(_4^{2-}\). The hydrogen evolution reaction at the cathode and oxygen evolution reaction at a higher current density might inhibit the oxidation of tartrazine during the AO/PS process, leading to stable color removal with a higher current density.

For the Pt-H\(_2\)O\(_2\)/PS process, at a reaction time of 20 min, PS concentration of 25%, and applied current densities of 6, 9, 12, and 15 mA cm\(^{-2}\), the color removals were obtained as 63%, 93%, 96%, and 97%, respectively, as shown in Figure 2c. Generally, a gradual increase in decolorization can be noticed with the increase of the current density from 6 to 12 mA cm\(^{-2}\). However, the color removal of tartrazine solution remained stable with the current density increased from 12 to 15 mA cm\(^{-2}\) in the Pt-H\(_2\)O\(_2\)/PS process and the BDD-H\(_2\)O\(_2\)/PS process. The abovementioned tendency can be attributed to the parasitic reactions at the anode and the cathode at high applied current density [32]. With the increase of current density, the hydrogen evolution reaction at the cathode and oxygen evolution reaction at the anode surface are enhanced, competing with the production of OH\(^-\) and SO\(_4^{2-}\). The hydrogen evolution reaction at the cathode and oxygen evolution reaction at a higher current density might inhibit the oxidation of tartrazine during the AO/PS process, leading to stable color removal with a higher current density.

For the BDD-H\(_2\)O\(_2\)/PS process, at a reaction time of 20 min, PS concentration of 25%, and applied current densities of 6, 9, 12, and 15 mA cm\(^{-2}\), the color removals were obtained as 53%, 95%, 96%, and 97%, respectively, as shown in Figure 2d. Generally, a gradual increase in decolorization can be noticed with the increase of the current density from 6 to 12 mA cm\(^{-2}\). However, the color removal of tartrazine solution remained stable with the current density increased from 12 to 15 mA cm\(^{-2}\) in the BDD-H\(_2\)O\(_2\)/PS process. The abovementioned tendency can be attributed to the parasitic reactions at the anode and the cathode at high applied current density [32]. With the increase of current density, the hydrogen evolution reaction at the cathode and oxygen evolution reaction at the anode surface are enhanced, competing with the production of OH\(^-\) and SO\(_4^{2-}\). The hydrogen evolution reaction at the cathode and oxygen evolution reaction at a higher current density might inhibit the oxidation of tartrazine during the AO/PS process, leading to stable color removal with a higher current density.

---

**Table 2. Effect of applied current density on color removal during AO/PS and AO-H\(_2\)O\(_2\)/PS processes.**

| Oxidation System | 6 mA cm\(^{-2}\) | 9 mA cm\(^{-2}\) | 12 mA cm\(^{-2}\) | 15 mA cm\(^{-2}\) |
|-----------------|-----------------|-----------------|-----------------|-----------------|
| Pt/PS           | 51              | 76              | 97              | 96              |
| BDD/PS          | 68              | 83              | 100             | 100             |
| Pt-H\(_2\)O\(_2\)/PS | 53              | 95              | 96              | 97              |
| BDD-H\(_2\)O\(_2\)/PS | 63              | 93              | 96              | 97              |

*For Pt/PS and BDD/PS, data of 20 min was selected. For Pt-H\(_2\)O\(_2\)/PS and BDD-H\(_2\)O\(_2\)/PS, data of 20 min was selected.*
tained as 53%, 95%, 96%, and 97%, respectively. Figure 2c highlights a sharp increase in decolorization in the Pt-H$_2$O$_2$/PS process with the increase of the current density from 6 to 9 mA cm$^{-2}$. This phenomenon may be attributed to the sharp increase in the accumulated concentration of H$_2$O$_2$, with the concentration of 19 mg L$^{-1}$ in the Pt-H$_2$O$_2$/PS process without tartrazine at a current density of 6 mA cm$^{-2}$ increased to 38 mg L$^{-1}$ at a current density of 9 mA cm$^{-2}$ at a reaction time of 45 min. However, further increase of current density to 12 mA cm$^{-2}$ merely caused a slight increase of H$_2$O$_2$, with accumulated concentration increased to 44 mg L$^{-1}$. With the increase of current density, the hydrogen evolution reaction at the cathode is promoted, inhibiting the electro-generation of H$_2$O$_2$ via Equation (4).

For the BDD-H$_2$O$_2$/PS process, at a reaction time of 20 min, a PS concentration of 50%, and applied current densities of 6, 9, 12, and 15 mA cm$^{-2}$, the color removals were obtained as 63%, 93%, 96%, and 97%, respectively, as shown in Figure 2d. The effect of current density in the BDD-H$_2$O$_2$/PS process presented an analogical tendency, as shown in the Pt-H$_2$O$_2$/PS process. The accumulated concentration of H$_2$O$_2$ in the BDD-H$_2$O$_2$/PS process without tartrazine at 6 mA cm$^{-2}$ is 13 mg L$^{-1}$, increasing to 34 mg L$^{-1}$ at a current density of 9 mA cm$^{-2}$ at a reaction time of 45 min. Further increase of current density to 12 mA cm$^{-2}$, the accumulated concentration only increased to 40 mg L$^{-1}$. Due to the strong oxidizing capacity of the BDD anode, electro-generated H$_2$O$_2$ can be oxidized by A-OH$^·$ at the BDD surface via Equation (13) [33]. Hence, the concentration of H$_2$O$_2$ in the BDD-H$_2$O$_2$/PS process is lower than that in the Pt-H$_2$O$_2$/PS process. Based on the above discussion, applied current density has a significant influence on color removal, and the optimum value for the AO/PS and AO-H$_2$O$_2$/PS processes was 12 mA cm$^{-2}$.

$$H_2O_2 + A-OH^· \rightarrow A-HO_2^· + H_2O \quad (13)$$

3.3. Effect of Reaction Temperature on Color Removal

Compared with other free radicals, PS can be activated by heat to generate SO$_4^{2−}$, a relatively low-cost and efficient activation method [34]. The effect of reaction temperature on color removal ranged from 25 to 45 $^\circ$C in the AO/PS process and the AO-H$_2$O$_2$/PS process, considering the tolerance and practicality of the experimental device. Table 3 shows the related degradation data during oxidation process.

| Oxidation System | 45 $^\circ$C | 45 $^\circ$C | 45 $^\circ$C |
|------------------|-------------|-------------|-------------|
| Pt/PS            | 80          | 84          | 90          |
| BDD/PS           | 94          | 97          | 98          |
| Pt-H$_2$O$_2$/PS | 80          | 81          | 82          |
| BDD-H$_2$O$_2$/PS | 90         | 91          | 93          |

* For Pt/PS and BDD/PS, data of 15 min was selected. For Pt-t-H$_2$O$_2$/PS and BDD-t-H$_2$O$_2$/PS, data of 15 min was selected.

For the Pt/PS process, at a reaction time of 15 min, PS concentration of 25%, and reaction temperature of 25, 35, and 45 $^\circ$C, the color removals were obtained as 80%, 84%, and 90%, respectively, as shown in Figure 3a. A slight increase in color removal can be attained with the increase of the reaction temperature, reflecting that the generation of SO$_4^{2−}$ activated by heat in the Pt/PS process occupied a relatively small proportion. Although higher temperatures can activate PS and enhance mass transfer during EAOPs, the critical factor affecting oxidation performance is the highly active oxidants absorbed at the anode surface [35]. A similar phenomenon was observed in the BDD/PS process, as shown in Figure 3b. At a reaction time of 15 min, PS concentration of 50%, and reaction temperatures of 25, 35, and 45 $^\circ$C, the color removals were obtained as 94%, 97%, and 98%,
respectively. Adding PS can increase the oxidation power of the AO/PS process at the Pt or BDD anodes. Still, the oxidation of tartrazine largely relies on the electrognerated oxidants absorbed at the anode surface; SO₄⁻ primarily comes from Equation (7) and can only promote color removal. From this discussion, one can deduce that the main source of SO₄⁻ in the AO/PS process is PS reduction at the cathode.

Figure 3. Cont.
Tert-butanol was used to scavenge A-OH processes, two well-known free radical scavengers, tert-butanol and methanol, were used. The application of tert-butanol and methanol, as shown in Figure 3c, exhibits a nearly similar loss of color in the Pt-H$_2$O$_2$/PS process with the increased reaction temperature. The slight increase in color removal may be caused by the almost unchanged concentration of H$_2$O$_2$ accumulated in the Pt-H$_2$O$_2$/PS process without tartrazine. At a reaction temperature of 25 °C, the concentration of H$_2$O$_2$ was 44 mg L$^{-1}$ after 45 min of electrolysis. Further increasing the temperature to 35 and 45 °C, the concentrations of H$_2$O$_2$ were 47 and 48 mg L$^{-1}$, respectively. This phenomenon may be attributed to the vital factors that affect the electrogenerated efficiency of H$_2$O$_2$. The key factors influencing the concentration of H$_2$O$_2$ in the AO-H$_2$O$_2$ and EF processes are the applied current density and the concentration of dissolved oxygen in the solution [36]. For the BDD-H$_2$O$_2$/PS process, the analogical tendency, as shown in Figure 3d, may also be caused by the concentration of H$_2$O$_2$ accumulated in solution without tartrazine, with the accumulated concentrations of H$_2$O$_2$ in the BDD-H$_2$O$_2$/PS process at reaction temperatures of 25, 35, and 45 °C were 40, 42, and 39 mg L$^{-1}$, respectively.

3.4. Oxidation Mechanism of Tartrazine in AO/PS and AO-H$_2$O$_2$/PS Process

The recipient standpoint of contaminations by the AO process in an aqueous solution is indirect oxidation, with the oxidation of contaminations initiated by reactive intermediates generated on the anode surface, including A-OH$^-$ and A=O. Furthermore, with the addition of PS, SO$_4^{2-}$ is introduced into the oxidation mechanism. To investigate the competitive generation of A-OH$^-$, OH$^-$, and SO$_4^{2-}$ in the AO/PS and AO-H$_2$O$_2$/PS processes, two well-known free radical scavengers, tert-butanol and methanol, were used. Tert-butanol was used to scavenge A-OH$^-$ and OH$^-$, whereas methanol was used to scavenge all these radicals [20]. For this purpose, all these radicals were exhaustively scavenged; the concentrations of tert-butanol and methanol are 500 times greater than that of tartrazine.

As shown in Figure 4a, the color removals were slightly affected by the addition of tert-butanol and methanol, indicating the low yields of A-OH$^-$, OH$^-$, and SO$_4^{2-}$ in the Pt/PS process. However, the loss of color removals in the BDD/PS process was more pronounced in the presence of tert-butanol and methanol, as shown in Figure 4b. The differences in color removals between the Pt/PS and BDD/PS processes with free radical scavengers may be ascribed to the property’s discrepancy of Pt and BDD anodes. The decolorization at the Pt anode is a mediated electrolysis by A=O and SO$_4^{2-}$ primarily comes from reduction reaction at the cathode via Equation (7). Nevertheless, the oxidation...
of tartrazine largely depends on the electrogenerated $A$-OH· at the BDD anode and $SO_4$·− generated through Equations (7) and (10).

Figure 4. Cont.
For the Pt-H₂O₂/PS process, a large decrease in final decolorization was observed compared to the Pt/PS process, as shown in Figure 4c. The decrease in oxidation power by the addition of free radical scavengers may cause by the activation reaction between PS and H₂O₂ as described in Equation (8), indicating that SO₄²⁻ and OH⁻ in the bulk of reaction solution promoted the color removal in the Pt-H₂O₂/PS process. Furthermore, the loss of color removals in the BDD-H₂O₂/PS process was more pronounced in the presence of tert-butanol and methanol, as shown in Figure 4d. According to this phenomenon, one can conclude that the color removal in the BDD-H₂O₂/PS process is dominated by these free radicals, especially A-·OH, HO₂·, and SO₄²⁻. The remaining color removal in the BDD-H₂O₂/PS process with radical scavengers may be caused by H₂O₂ and A-HO₂·.

Based on the above analysis, these oxidizing agents played an important role in the BDD/PS, Pt-H₂O₂/PS, and BDD-H₂O₂/PS processes. During the Pt/PS process, the decolorization is caused by A=O and SO₄²⁻, primarily coming from the reduction reaction at the cathode via Equation (7). In the BDD/PS process, the oxidation of tartrazine largely depends on the electrogenerated A-·OH at the BDD anode and SO₄²⁻ generated through Equations (7) and (10). However, for the AO-H₂O₂/PS process, adding PS is an effective strategy to improve color removal since the generations of SO₄²⁻ and OH⁻ rely on Equation (8) to introduce large numbers of free radicals into the system. Based on the above analysis, the promotive effect of PS at Pt and BDD anodes during the AO and AO-H₂O₂ processes is proposed in Figure 5.

To confirm the effect of dissolved oxygen in the AO-H₂O₂/PS system, N₂ was passed into the reaction solution at a flow rate of 1 L min⁻¹ during the Pt-H₂O₂/PS process and in the BDD-H₂O₂/PS process instead of the air. For the Pt-H₂O₂/PS process, at a reaction time of 15 min, PS concentration of 25%, applied current density: 12 mA cm⁻², and reaction temperature of 25 °C, the color removals were obtained as 75%, lower than that of the Pt/PS process. For the BDD-H₂O₂/PS process, the same phenomenon was noticed, as the color removals were only 79% lower than that of the BDD/PS process. N₂ experiments proved that dissolved oxygen promoted the degradation of tartrazine in the AO-H₂O₂/PS system.
Figure 5. Proposed electrochemical oxidation mechanism in AO/PS and AO-H$_2$O$_2$/PS process.

Tartrazine can be completely removed in AO/PS and AO-H$_2$O$_2$/PS system after 45 min oxidation. However, the decoloration process of dyes by AO and related technologies is usually detected from the abatement of the COD values. To evaluate the oxidation ability of the AO/PS and AO-H$_2$O$_2$/PS system on the COD, the oxidation experiments in aqueous solutions containing 250 mg L$^{-1}$ tartrazine under an applied current density of 12 mA cm$^{-2}$ and initial pH value of 6.8 were conducted. The variations of COD values with different oxidation systems are shown in Figure 6. For the AO/PS process, at a reaction time of 45 min, the COD removals were obtained as 23% for the Pt and 26% for the BDD anode. For the AO-H$_2$O$_2$/PS process, at a reaction time of 45 min, the COD removals were obtained as 34% for the Pt and 38% for the BDD anode. Hence, we could deduce that tartrazine was more effectively degraded and oxidized during the AO process than oxidation products.

Figure 6. COD removal efficiency in AO system. Raw dye solution: 250 mg L$^{-1}$ tartrazine, initial pH: 6.8, reaction temperature: 25 °C, applied current density: 12 mA cm$^{-2}$.

3.5. Economic Analysis

During EAOPs, it was clear that the main cost was power and electrode consumption. Considering the energy and PS consumption, the direct cost of tartrazine treatment by the BDD-H$_2$O$_2$/PS process was calculated. The electricity cost for treating 1 L of tartrazine solutions by the BDD-H$_2$O$_2$/PS process under the optimum conditions was $0.0002. The PS cost for treating 1 L of tartrazine solutions by the BDD-H$_2$O$_2$/PS process under the
optimum conditions was $0.0051. Hence, the oxidation cost of tartrazine by BDD-H$_2$O$_2$/PS was $0.0053 per L. Other oxidation methods for treatment of refractory organics have been reported, such as coagulation cascaded with granular activated carbon, O$_3$/H$_2$O$_2$ and solar photo-Fenton, with the corresponding treatment costs of $0.0166, $0.0125, and $0.0147 [37] per L, respectively. Compared with traditional catalytic oxidation technologies, EAOPs hardly require a catalyst, and the electrode lifetime is much longer than that of a catalyst. Therefore, the cost of an electrode can be reduced by the catalyst cost and the usage of alternate forms of energy (e.g., solar energy and wind energy), and to confirm the stability of the AO/PS and AO-H$_2$O$_2$/PS system, 5 consecutive cycles experiments were conducted. Thanks to the stability of commercial anodes (Pt and BDD), no decreasing signs of dye degradation efficiency were observed, indicating that the AO/PS and AO-H$_2$O$_2$/PS systems have good stability and meet the requirement of long-term continuous operation.

4. Conclusions

It has been demonstrated that activating PS by electrogenerated H$_2$O$_2$ and anodic oxidation promotes the decolorization of tartrazine solutions at the Pt and BDD anodes. For the Pt anode, SO$_4^{2-}$ was generated from the reduction reaction at the cathode and reaction between PS and H$_2$O$_2$ with the optimal concentration of PS at 25%. For the BDD anode, SO$_4^{2-}$ was generated from the reduction reaction at the cathode, oxidation reaction at the anode, and reaction between PS and H$_2$O$_2$ with the optimal concentration of PS at 50%. Among these activation methods, the activation efficiency of PS by H$_2$O$_2$ is much better than other methods. The processes become more efficient based on the increase of the applied current density to 12 mA cm$^{-2}$ due to the more production of SO$_4^{2-}$ and OH· under high current density. The PS activations by thermal show much the same performance with the temperature range from 25 to 45°C. The main oxidizing agents are SO$_4^{2-}$ and OH· in AO/PS and AO-H$_2$O$_2$/PS, and the yields of these radicals in the AO/PS process are relatively low.

Author Contributions: K.Z.: Supervision, Resources, and Writing original draft. Y.Y. and Y.L.: Investigation, Validation, and Data analysis. Q.L.: Project administration, Methodology, Data curation. L.H.: Writing-review & editing. All authors have read and agreed to the published version of the manuscript.

Funding: This work was financially supported by the Natural Science Foundation of Shandong Province (ZR2021ME187) and the Project of Introducing and Cultivating Young Talent in the Universities of Shandong Province (Soil Erosion Process and Ecological Regulation).

Institutional Review Board Statement: Not applicable.

Informed Consent Statement: Not applicable.

Conflicts of Interest: The authors declare no conflict of interest.

References

1. Alderete, B.L.; da Silva, J.; Godoi, R.; da Silva, F.R.; Taffarel, S.R.; da Silva, L.P.; Garcia, A.L.H.; Júnior, H.M.; de Amorim, H.L.N.; Picada, J.N. Evaluation of toxicity and mutagenicity of a synthetic effluent containing azo dye after Advanced Oxidation Process treatment. Chemosphere 2020, 263, 128291. [CrossRef] [PubMed]
2. Badvi, K.; Javanbakht, V. Enhanced photocatalytic degradation of dye contaminants with TiO$_2$ immobilized on ZSM-5 zeolite modified with nickel nanoparticles. J. Clean. Prod. 2021, 280, 124518. [CrossRef]
3. Balçık, U.; Chormey, D.S.; Ayyıldız, M.F.; Bakirdere, S. Liquid phase microextraction based sensitive analytical strategy for the determination of 22 hazardous aromatic amine products of azo dyes in wastewater and tap water samples by GC-MS system. Microchem. J. 2020, 155, 104712. [CrossRef]
4. Hu, E.; Wu, X.; Shang, S.; Tao, X.; Jiang, S.; Gan, L. Catalytic ozonation of simulated textile dyeing wastewater using mesoporous carbon aerogel supported copper oxide catalyst. J. Clean. Prod. 2016, 112, 4710–4718. [CrossRef]
5. Xia, Y.; Wang, G.; Guo, L.; Dai, Q.; Ma, X. Electrochemical oxidation of Acid Orange 7 azo dye using a PbO$_2$ electrode: Parameter optimization, reaction mechanism and toxicity evaluation. Chemosphere 2020, 241, 125010. [CrossRef]
6. Wang, J.; Yao, J.; Wang, L.; Xue, Q.; Hu, Z.; Pan, B. Multivariate optimization of the pulse electrochemical oxidation for treating recalcitrant dye wastewater. Sep. Purif. Technol. 2020, 230, 115851. [CrossRef]
7. Zhang, J.; Wei, X.; Miao, J.; Zhang, R.; Zhang, J.; Zhou, M.; Lu, W. Enhanced performance of an Al-doped SnO$_2$ anode for the electrocatalytic oxidation of organic pollutants in water. *Mater. Today Commun.* 2020, 24, 101164. [CrossRef]

8. Zhu, K.; Qi, H.; Sun, X.; Sun, Z. Anodic oxidation of diuron using Co$_3$O$_4$/graphite composite electrode at low applied current. *Electrochim. Acta* 2019, 299, 853–862. [CrossRef]

9. Tang, Y.; He, D.; Guo, Y.; Qu, W.; Shang, J.; Zhou, L.; Pan, R.; Dong, W. Electrochemical oxidative degradation of X-6G dye by boron-doped diamond anodes: Effect of operating parameters. *Chemosphere* 2020, 258, 127368. [CrossRef]

10. Rueffer, M.; Bejan, D.; Bunce, N.J. Graphite: An active or an inactive anode? *Electrochim. Acta* 2011, 56, 2246–2253. [CrossRef]

11. Jiang, Y.; Zhao, H.; Liang, J.; Yue, L.; Li, T.; Luo, Y.; Liu, Q.; Lu, S.; Asiri, A.M.; Gong, Z.; et al. Anodic oxidation for the degradation of organic pollutants: Anode materials, operating conditions and mechanisms. A mini review. *Electrochem. Commun.* 2021, 123, 106912. [CrossRef]

12. Wang, J.; Li, C.; Rauf, M.; Luo, H.; Sun, X.; Jiang, Y. Gas diffusion electrodes for H$_2$O$_2$ production and their applications for electrochemical degradation of organic pollutants in water: A review. *Sci. Total Environ.* 2020, 759, 143459. [CrossRef] [PubMed]

13. Martínez-Huitle, C.A.; Rodrigo, M.A.; Sirés, I.; Scialdone, O. Single and Coupled Electrochemical Processes and Reactors for the Abatement of Organic Water Pollutants: A Critical Review. *Chem. Rev.* 2015, 115, 13362–13407. [CrossRef] [PubMed]

14. Li, D.; Zheng, T.; Liu, Y.; Hou, D.; He, H.; Song, H.; Zhang, J.; Tian, S.; Zhang, W.; Wang, L.; et al. A cost-effective Electro-Fenton process with graphite felt electrode aeration for degradation of dimethyl phthalate: Enhanced generation of H$_2$O$_2$ and iron recycling that simultaneously regenerates the electrode. *Chem. Eng. J.* 2020, 394, 125033. [CrossRef]

15. Qi, C.; Yu, G.; Huang, J.; Wang, B.; Wang, Y.; Deng, S. Activation of persulfate by modified drinking water treatment residuals for sulfamethoxazole degradation. *Chem. Eng. J.* 2018, 335, 490–498. [CrossRef]

16. Ren, T.-L.; Ma, X.-W.; Wu, X.-Q.; Yuan, L.; Lai, Y.-L.; Tong, Z.-H. Degradation of imidazolium ionic liquids in a thermally activated persulfate system. *Chem. Eng. J.* 2021, 412, 128624. [CrossRef]

17. Devi, P.; Das, U.; Dalai, A.K. In-situ chemical oxidation: Principle and applications of peroxide and persulfate treatments in wastewater systems. *Sci. Total Environ.* 2016, 571, 643–657. [CrossRef] [PubMed]

18. Luo, H.; Cheng, Y.; Zeng, Y.; Luo, K.; He, D.; Pan, X. Rapid removal of organic micropollutants by heterogeneous peroxymono-sulfate catalysis over a wide pH range: Performance, mechanism and economic analysis. *Sep. Purif. Technol.* 2020, 248, 117023. [CrossRef]

19. Matzek, L.W.; Carter, K.E. Activated persulfate for organic chemical degradation: A review. *Chemosphere* 2016, 151, 178–188. [CrossRef]

20. Dos Santos, A.J.; Brillas, E.; Cabot, P.L.; Sirés, I. Simultaneous persulfate activation by electrogenerated H$_2$O$_2$ and anodic oxidation at a boron-doped diamond anode for the treatment of dye solutions. *Sci. Total Environ.* 2020, 747, 141541. [CrossRef]

21. Wu, J.; Wang, B.; Cagnetta, G.; Huang, J.; Wang, Y.; Deng, S.; Yu, G. Nanoscale zero valent iron-activated persulfate coupled with Fenton oxidation process for typical pharmaceuticals and personal care products degradation. *Sep. Purif. Technol.* 2020, 239, 116534. [CrossRef]

22. Jiang, C.; Yang, Y.; Zhang, L.; Lu, D.; Lu, L.; Yang, X.; Cai, T. Degradation of Atrazine, Simazine and Ametryn in an arable soil using thermal-activated persulfate oxidation process: Optimization, kinetics, and degradation pathway. *J. Hazard. Mater.* 2020, 400, 123201. [CrossRef]

23. Silveira, J.E.; Garcia-Costa, A.L.; Cardoso, T.O.; Zazo, J.A.; Casas, J.A. Indirect decolorization of azo dye Disperse Blue 3 by electro-activated persulfate. *Electrochim. Acta* 2017, 258, 927–932. [CrossRef]

24. Zhang, W.; He, Y.; Li, C.; Hu, X.; Yang, S.; You, X.; Liang, W. Persulfate activation using Co/AC particle electrodes and synergetic effects on hemic acid degradation. *Appl. Catal. B Environ.* 2021, 285, 119848. [CrossRef]

25. Cai, J.; Zhou, M.; Du, X.; Xu, X. Enhanced mechanism of 2,4-dichlorophenoxyacetic acid degradation by electrochemical activation of persulfate on Blue-TiO$_2$ nanotubes anode. *Sep. Purif. Technol.* 2021, 254, 117560. [CrossRef]

26. Ukundimana, Z.; Omwene, P.; Gengec, E.; Can, O.; Kobya, M. Electrooxidation as post treatment of ultrafiltration effluent in a landfill leachate MBR treatment plant: Effects of BDD, Pt and DSA anode types. *Electrochim. Acta* 2019, 285, 141541. [CrossRef]

27. El-Ghenemy, A.; Centellas, F.; Rodríguez-Huitle, C.A.; Rodríguez, M.; Casas, J.A.; Sirés, I.; Brillas, E. Comparative use of anodic oxidation, electro-Fenton and photoelectro-Fenton with Pt or boron-doped diamond anode to decolorize and mineralize Malachite Green oxalate dye. *Electrochim. Acta* 2015, 182, 247–256. [CrossRef]

28. Cai, J.; Zhou, M.; Pan, Y.; Lu, X. Degradation of 2,4-dichlorophenoxyacetic acid by anodic oxidation and electro-Fenton using BDD anode: Influencing factors and mechanism. *Sep. Purif. Technol.* 2019, 230, 115867. [CrossRef]

29. Clematis, D.; Panizza, M. Electro-Fenton, solar photoelectro-Fenton and UVA photoelectro-Fenton: Degradation of Erythrosine B dye solution. *Chemom.2020, 270, 129480. [CrossRef]

30. Feng, H.; Chen, Z.; Wang, X.; Chen, S.; Cattenden, J. Electrochemical advanced oxidation for treating ultrafiltration effluent of a landfill leachate system: Impacts of organics and inorganics and economic evaluation. *Chem. Eng. J.* 2020, 413, 127492. [CrossRef]

31. Chankiya, P.; Nidheesh, P.; Babu, D.S.; Gopinath, A.; Kumar, M.S. Treatment of dyeing wastewater by combined sulfide radical based electrochemical advanced oxidation and electrocoagulation processes. *Sep. Purif. Technol.* 2020, 254, 117570. [CrossRef]

32. Wu, P.; Zhang, Y.; Chen, Z.; Duan, Y.; Lai, Y.; Fang, Q.; Wang, F.; Li, S. Performance of boron-doped graphene aerogel modified gas diffusion electrode for in-situ metal-free electrochemical advanced oxidation of Bisphenol A. *Appl. Catal. B Environ.* 2019, 255, 117784. [CrossRef]
33. Brillas, E.; Sirés, I.; Oturan, M.A. Electro-Fenton Process and Related Electrochemical Technologies Based on Fenton’s Reaction Chemistry. *Chem. Rev.* 2009, 109, 6570–6631. [CrossRef]

34. Huang, D.; Chen, N.; Zhu, C.; Fang, G.; Zhou, D. The overlooked oxidative dissolution of silver sulfide nanoparticles by thermal activation of persulfate: Processes, mechanisms, and influencing factors. *Sci. Total Environ.* 2020, 760, 144504. [CrossRef]

35. Isarain-Chávez, E.; Baró, M.D.; Rossinyol, E.; Morales-Ortiz, U.; Sort, J.; Brillas, E.; Pellicer, E. Comparative electrochemical oxidation of methyl orange azo dye using Ti/Ir-Pb, Ti/Ir-Sn, Ti/Ru-Pb, Ti/Pt-Pd and Ti/RuO₂ anodes. *Electrochim. Acta* 2017, 244, 199–208. [CrossRef]

36. Yang, B.; Wei, T.; Xiao, K.; Deng, J.; Yu, G.; Deng, S.; Li, J.; Zhu, C.; Duan, H.; Zhuo, Q. Effective mineralization of anti-epilepsy drug carbamazepine in aqueous solution by simultaneously electro-generated H₂O₂/O₃ process. *Electrochim. Acta* 2018, 290, 203–210. [CrossRef]

37. Deng, Y.; Chen, N.; Feng, C.; Chen, F.; Wang, H.; Kuang, P.; Feng, Z.; Liu, T.; Gao, Y.; Hu, W. Treatment of organic wastewater containing nitrogen and chlorine by combinatorial electrochemical system: Taking biologically treated landfill leachate treatment as an example. *Chem. Eng. J.* 2019, 364, 349–360. [CrossRef]