Pesticides in Worldwide Aquatic Systems: Part II

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Abstract

Contamination by pesticides is a worldwide problem that can greatly disturb the biota, directly and/or indirectly. Nonetheless, few efforts were done so far to present review-style publications that analyse and integrate monitoring data—in a global scale—and evaluate possible environmental risks. Herein, we assessed possible environmental risks through theoretical calculations, using worldwide data published at least during the last 17 years and considering different trophic levels and the maximum average environmental concentrations (in water) observed in each continent. Furthermore, hazard quotients—using the estimated average daily intake, theoretical maximum daily intake and the maximum residue limits—were calculated to estimate the potential risks to humans through direct consumption of molluscs, crustaceans and fish. In summary, several pesticides were quantified at concentrations capable to affect low to medium trophic level species, which through the food web can affect higher trophic levels; theoretical approaches considering the environmental mixtures showed that algae and invertebrates are the most sensitive groups. Moreover, fish and crustaceans evidenced the highest body concentrations. To evaluate a potential risk through direct consumption, human health risk assessments were done, and in spite of no direct risk, some hazard quotients indicate a potential risk for developing carcinogenic effects.

Keywords: insecticides, herbicides, fungicides, aquatic organisms, EC$_{50}$, LC$_{50}$, PNEC, ADI, EADI, MRL, hazard quotients, mollusc, crustacean, bivalve, fish, bioaccumulation, biomagnification
1. Preamble

Worldwide, several studies have shown contamination with pesticides within different matrices. Together with the data shown previously in chapter “Pesticides in Worldwide Aquatic Systems: Part I”, information such as the maximum concentrations in waters and the concentration of pesticides found in the different biological matrices were used to (i) assess eventual individual pesticide risk through the comparison with the well-established EC$_{50}$/LC$_{50}$ for aquatic organisms, (ii) predict the environmental risk from pesticide mixtures found in each continent and (iii) assess the potential risk for human health when consuming molluscs, crustaceans and fish with the quantified concentrations.

2. Aquatic organisms

Fifty-two studies were used, where 111 different species were studied. The continent with the highest percentage of available results (quantified pesticides in different organisms) is Africa (39 species), followed by Europe (35 species), Asia (26 species) and then North and South America (nine and eight species, respectively). Here, we decided to focus on the sample type (zooplankton, molluscs, crustaceans, fish and mammals) analysed per continent and country (Table 1).

| Continent/country | Number of aquatic systems | Quantified pesticides | Sampling year | Sample type | av-min | av-max | av-av | References |
|-------------------|---------------------------|-----------------------|---------------|-------------|--------|--------|--------|------------|
| Africa            |                           |                       |               |             |        |        |        |            |
| Egypt             | 2                         | 14                    | 1993          | C, F        | 1.1–6.3| 7.6–8.2| 4.1–130.2| [3]        |
| Ethiopia          | 1                         | 12                    | 2011          | F           | 4.1    | 27.2   | 17.1   | [4]        |
| Ghana             | 3                         | 6–13                  | 2004–2015     | F           | 1.6–79.8| 2.8–154.3| 1.6–120.5| [4, 5]     |
| Kenya             | 1                         | 7                     | 2011          | F           | na     | na     | 0.3    | [1, 6, 7]  |
| Nigeria           | 5                         | 1–65                  | 2003–2014     | F           | 19.5–3618| 21.5–6355| 20.6–5233|[1, 7–9]    |
| Tunisia           | 1                         | 14                    | 2010          | F           | 14.9   | 39.3   | 22.1   | [10]       |
| Asia              |                           |                       |               |             |        |        |        |            |
| China             | 5                         | 7–45                  | 2003–2013     | F, Mo, C    | 0.6–2.5| 1.8–34.5| 0.8–11.8| [11–15]    |
| India             | 1                         | 3                     | na            | Ma          | na     | na     | 74.5   | [16]       |
| Russia            | 8                         | 2012–2013             | F             | 18.1        | 52.1   | 31.7   |        | [17]       |
| South Korea       | 1                         | 15                    | na            | F           | na     | na     | 2.2    | [18]       |
| Tibet             | 1                         | 55                    | 2005          | F           | na     | na     | 1.0    | [19]       |
| Europe            |                           |                       |               |             |        |        |        |            |
| Baltic Sea        | 1                         | 18                    | 2003          | C, F, Mo    | 8.1    | 10.8   | 9.4    | [20]       |
| Belgium           | 1                         | 5                     | 2001          | F, Mo       | 1.5    | 7.6    | 4.1    | [21, 22]   |
The data collected between 1993 and 2016 averaged from 0.1 to 5233 ng/g (Table 1). Europe is represented by 22 aquatic systems, followed by Africa, with 13, and the rest with no more than nine aquatic systems. When considering the number of pesticides quantified, Africa has more observations (382) than Europe (327) and the other continents (between 90 and 220), which is due to the higher number of species studied in Africa.

Africa stands out with average concentrations of 132 ng/g (SD 411), followed by Europe (57 ng/g, SD 271), South America (17 ng/g, SD 40) and Asia and North America (5 ng/g, SD 20). This scattered difference between concentrations is mainly due to the average values observed in Warri River (Nigeria, Africa) and in the Danube Delta (Romania, Europe) [1, 2].

Grouping data by category, insecticides prevail in 89% of biologic analyses, leaving 11% for the herbicide and fungicide categories and presenting the same pattern on all continents (Figure 1).
Figure 1. Representation of the quantified pesticides in organisms (%), per category, in each continent; the right upper corner pie chart represents the Metazoan lineages used worldwide.

Figure 2. Representation of the quantified pesticides in organisms (%), per lineages of Metazoan, vertebrates and invertebrates and matrices.
Figure 3. Average pesticide concentrations (ng/g fresh weigh) and number of quantifications per Metazoan lineage, worldwide (A) and by continent (B). The error bars represent standard deviations (SD).
No data are available for Oceania and Antarctica; so, when citing herein “worldwide”, these continents will not appear.

Analysing data by matrix, the most common analysis occurred in fish (74%) and molluscs (20%). The remaining studies considered zooplankton, crustaceans, turtles and mammals (Figure 1). In total, 74% of the quantified pesticides were conducted on vertebrates and the other 26% in invertebrates (Figure 1). While for the latter, 80% of the quantifications were done using the whole animal, and for vertebrates, it is further divided; specific organs or tissues were used to quantify pesticides.

Many factors account for the strong bias towards vertebrates. Invertebrates are small, less complex and as a food resource almost entirely eatable, while the same is not applicable to vertebrates. Besides that, the study goal (i.e., food control or environmental/toxicological studies) also influences the type of the tissue/organ to study (muscle, liver, gonads or gills). For example, the bubbler tissue and fat are only applicable for aquatic mammals and turtles (Figure 2).

Results per Metazoan lineages (zooplankton, mollusc, crustacean, fish, turtles and mammals) were assessed considering the average concentrations and the number of quantifications (Figure 3). Average concentrations were ~11 ng/g for zooplankton and molluscs, ~35 ng/g for mammals and ~100 ng/g for crustaceans and fishes (Figure 3A).

Among continents, Africa presented the highest concentrations for crustaceans (142 ng/g) and fishes (253 ng/g) followed by North America (136 ng/g for crustaceans). Asia, Europe and South America included data belonging to four Metazoan lineages, with similar range of concentrations (~3 to ~76 ng/g (Figure 3B)).

3. Half effective and lethal concentrations (EC50/LC50) for aquatic organisms

It is now well established that at specific concentrations all pesticides are harmful to biota, affecting algae and plants, invertebrates and vertebrates [51]. Databases such as Pesticides Properties DataBase (PPDB) present information on the physicochemical properties, environmental fate, human health and ecotoxicological data of all active ingredients and approved pesticides [52].

In order to evaluate the worst-case scenario, the maximum average concentrations measured in waters from each continent were compared against the acute and chronic concentrations for aquatic animals, documented by the PPDB (see chapter “Part I: Pesticides in Worldwide Aquatic Systems”). On a global scale, 57 pesticides were registered at maximum average concentrations above the LC50 and/or EC50 settled for algae, invertebrates and/or fishes; among continents, Europe reported the highest number of pesticides (44 of 116), followed by Asia (14 of 42), Africa, Oceania and finally South America (6 of 24).

The most critical measured environmental concentrations (MEC) were registered for dicofol, ethion (Asia), metribuzin (Europe) and diazinon (Africa) with values from 2- to 200-folds higher than EC50 or LC50 set for invertebrates and algae.
4. Predictive aquatic risk assessment of pesticide mixtures

Despite the common occurrence of pesticides, mixtures in the environment, laws, conventions and recommendations still focus on individual standard parameters. Modelling approaches, based on available ecotoxicological information, can be used to estimate the impact of mixtures in the biota, completing this lack of information [53].

Based on the European chemical legislation REACH, the ecological risk quotient (RQ) is determined by the equation:

\[
RQ = \frac{\text{MEC}}{\text{PNEC}} = \frac{\text{Measured Environmental Concentration (MEC; mg/L)}}{\text{Predicted No Effect Concentration (PNEC; mg/L)}}
\]

PNEC is derived by selecting the most sensitive trophic level—from algae, invertebrate or fish—and applying an appropriate assessment factor (AF) [52, 54]. The AF, also denoted as safety or uncertainty factor, considers intra- and inter-laboratory variation of the data, biological variance and short-term to long-term exposures, presenting stipulated values for specific conditions [55, 56]; as an example, considering the Maximum Acceptable Concentration-Quality Standards (MAC-QS) to assess short-term effects an AF = 100 should be applied [55].

The RQ values, classified from <0.01 (negligible) to >1 (significant), indicate a range of potential risks for concern, but do not inform about the specific biological end point for that organism which is representing a specific trophic level [53, 57]. For this reason, a second approach, which defines the most sensitive trophic level for that environmental concentration, should be applied [53]:

\[
RQ_{\text{toxic units (TU)}} = \frac{\text{MEC (mg/L)}}{\text{EC}_{50} \text{ or } \text{LC}_{50} \text{ per each trophic level (mg/L)}}
\]

RQ_{TU} values are summed per trophic level (sum of the toxic units (RQ_{STU})). If both RQ_{(MEC/PNEC)} and RQ_{STU} are >1, additional considerations are required [53]. Based on the two reference models—concentration addition (CA) and independent action (IA)—the RQ_{STU}/Max_{TU} can be used to predict the second tier, resulting in the maximum value from which CA may display higher toxicity values than IA [58].

In this work, the maximum of the average measured concentration of pesticides in water samples was used to assess the potential risk per continent and on a worldwide scale (Table 2). From a total of 144 pesticides quantified in water samples, 133 were used for ecological risk assessment (Table 2); the remainders, mostly isomers and metabolites, were not integrated due to lack of information on their EC_{50} and LC_{50} concentrations set for these trophic levels (algae, invertebrate and fish). The highest number of pesticides suitable for this approach are represented by insecticides (n = 118). In general, algae was the most sensitive group to herbicides and fungicides, with 75% and 61.5% of the cases, respectively, while invertebrates showed the highest sensitivity to insecticides (66.1%) (Table 2).

Globally, the RQ_{(MEC/PNEC)} resulted in 43% of very high-risk cases, led by insecticides; fungicides were the least worrisome category, as most of the cases presented low or negligible risks (Figure 4).
The results presented above are a consequence of the highest values measured around the world. Since Europe was the continent with more values of $RQ(MEC/PNEC)$, these results are mostly representative for this continent (see Table 1). However, this does not mean that concentrations measured on the other continents are innocuous. As observed for the number of compounds analysed per continent, Africa presented the most disturbing scenarios (52%), followed by Asia and Europe (45%) and then Oceania and South America (24%) with $RQ > 1$ (Figure 5).

In order to evaluate the effect of the maximum average concentrations found per individual trophic level ($RQ_{TU}$), further evaluation should be done through $RQ_{STU}$ (Table 3).

When comparing between continents, the highest $RQ_{STU}$ ratios were attained in Europe, for algae (16.13) and fish (33.12), and in Asia for invertebrates (324.97); however, the last one is due to a punctual concentration observed in India for ethion [59]. Independently of that, the invertebrate group is the most sensitive trophic level, presenting the highest $RQ_{STU}$ values. The same pattern is observed in the other continents except in Oceania, where the highest risk is observed for the algae (0.92) by the herbicides (Table 3).

The $RQ(MEC/PNEC)$ and $RQ_{STU}$ demonstrate that one or more biotest organisms are sensitive to the concentrations presented on that continent; so, the ratio $RQ_{STU}/$highest $RQ_{TU}$ was done, applying the highest sum among trophic levels (Table 4).

For each of these scenarios, the maximal possible ratio $RQ_{STU}/RQ_{TU}$ was lower than the value given by the number of mixture of toxic components, suggesting that the possible observed toxicity is due to a low number of pesticides. As we can notice, the $RQ_{STU}/RQ_{TU}$ ratio is very

|         | Africa | Asia  | Europe | Oceania | South America | PNEC | Algae | Invert | Fish |
|---------|--------|-------|--------|---------|--------------|------|-------|--------|------|
|         | Av.    |       | Av.    | Av.     | Av.          |      |       |        |      |
|         | mg/L   | mg/L  | mg/L   | mg/L    | mg/L         | (mg/L) | %     |        |      |
| Herbicides | 5.7E-04 | 8.1E-05 | 7.8E-03 | 5.7E-04 | 2.3E-03 | 2.5E-06–2.0E+00 | 75.0  | 6.25  | 18.8  |
| n       | 5      | 6     | 38     | 9       | 7            |       |       |        |      |
| Insecticides | 7.7E-04 | 2.2E-03 | 2.3E-03 | 1.4E-05 | 2.9E-04 | 1.7E-08–1.0E+00 | 8.5   | 25.4  | 66.1  |
| n       | 19     | 28    | 49     | 7       | 15           |       |       |        |      |
| Fungicides  | 8.5E-05 | 4.5E-04 | 2.6E-04 | 1.1E-05 | 3.9E-05 | 3.0E-05–4.6E-01 | 61.5  | 0.0   | 38.5  |
| n       | 1      | 8     | 22     | 3       | 3            |       |       |        |      |

Table 2. Ecological risk assessment through the PNEC, using the maximum average concentrations of pesticides in water (mg/L), quantified in each continent; here in this table only the average values/category (Av.), the total number of pesticides (n) observed per category and continent, and the range of PNEC values are presented; data based on Table 1 of the chapter in this book entitled Pesticides in worldwide aquatic systems- Part I.
Figure 4. Distribution of pesticides per category (%), according to $\text{RQ}_{\text{MEC/PNEC}}$ ranking.

Figure 5. Percentage of $\text{RQ}_{\text{MEC/PNEC}}$ samples above 1, grouped by continent (total number of observations, n = 25, 42, 110, 19, 25 in Africa, Asia, Europe, Oceania, and South America, respectively).
similar among continents; however, the number of toxic compounds per total, where Africa presents a significant number (52%) when compared to the others, should be also considered (Table 4).

### Table 4. Second tier, using RQSTU and the highest RQTU per trophic level and continent.

| Continent     | No. of compounds (toxic/total) | RQTU   | \(\Sigma RQ_{STU}\) | \(\Sigma RQ_{STU}/RQTU\) |
|---------------|--------------------------------|--------|-----------------------|---------------------------|
|                |                                |        | Algae                 | Invert                    | Fishes                    |
| Africa        | 13/25                          | Parathion methyl | 8.05 | 9.22 | 1.15 |
| Asia          | 10/21                          | Ethion  | 221.42 | 8.45 | 1.47 |
| Europe        | 9/22                           | Deltamethrin | 20.11 | 33.12 | 1.65 |
| Oceania       | 5/19                           | Diuron  | 0.79 | 0.92 | 1.17 |
| South America | 6/25                           | Cypermethrin | 2.47 | 3.96 | 1.61 |

### 5. Human health risks

Dietary pesticide risks can be estimated through well-established indices and defined and used by US Environmental Protection Agency (EPA) [60], European Food Safety Authority (EFSA)
and European Union Directives [61, 62]. Realistic predictions involve several parameters, such as pesticide residue intake (PRI), as the one reported by Food and Agriculture Organisation (FAO) [63]:

\[
PRI = \text{Pesticide Concentration (mg/kg)} \times \text{Acceptable Consumption Rate (kg/capita/day)}
\]

The acceptable daily intake (ADI) estimates the amount of a substance in food that can be ingested daily over a lifetime without appreciable health risk to the consumer [64]:

\[
ADI(\text{mg/kg/day}) = \frac{\text{No observed Effect Level (NOEL)}}{\text{Safety Factor}}
\]

The estimated average daily intake (EADI), according to EPA, should be less than the established ADI values [64]:

\[
EADI(\text{mg kg bw/day}) = \frac{\text{PRI}}{\text{Standard Body Weight}}
\]

The theoretical maximum daily intake (TMDI) represents the maximum concentration of a pesticide residue (mg/kg) legally permitted in food [64]:

\[
\text{TMDI} = \text{Consumption Rate (kg/capita/day)} \times \text{Maximum Residue Limits (MRLs)}
\]

When no specific MRL is published, a 0.01 mg/kg value is applied [8]. Additionally, hazard quotients (HQs)—which measure the potential exposure for developing non-carcinogenic health effects—may be calculated using several assumptions [65].

EADI may be divided by the acute reference dose (ARfD, mg/kg/day) [14]—which is derived from the no-observed-adverse-effect levels (NOAEL) and based on studies of short time exposures (1–7 days) [66], by ADI, for long intake periods, and by TMDI, which is advised by EFSA to calculate the potential risks of unintentional compounds, such as pollutants.

In the chapter Part I: Pesticides in Worldwide Aquatic Systems, the levels/categories of pesticides per continent/country are displayed. The maximum average concentrations shown in Part I were used here to assess human health risks. Data are summarised in Table 5.

| Continent     | Molluscs  | Invertebrates | Fishes     |
|---------------|-----------|---------------|------------|
| Africa        | —         | 1.6E + 00 (30)| 5.2E + 00 (370) |
| Asia          | 3.2E-01 (41)| 1.0E-02 (3)| 9.0E-02 (173) |
| Europe        | 6.0E-02 (143)| 1.4E-01 (10)| 4.0E + 00 (168) |
| North America | 1.0E-02 (16)| 1.4E-01 (1)| 3.0E-02 (104) |
| South America | 1.8E-01 (46)| —             | 1.7E-01 (32) |

Table 5. Average maximum concentrations (mg/kg) found per continent and by group of aquatic animals (mollusc, invertebrates and fish) and the total number of cases used in each case (between brackets).
| Continent/pesticide | Mollusc | Asia | HCH (gamma) | 2.9E-03 | 2.9E-06 | 2.0E-02 | 1.1E-03 | 8.0E-03 | 3.0E-04 | 0 | 0.02 | 0.08 | 0 |
| --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- |
| HCH (sigma) | 3.2E-01 | 3.1E-04 | 1.0E-02 | 5.7E-04 | — | — | 0.03 | 0.09 | 0.21 | — | — |
| Heptachlor epoxide | 5.8E-04 | 5.8E-07 | 4.0E-03 | 2.3E-04 | 1.0E-04 | — | 0 | 0.07 | 0.08 | — | — |
| Methoxychlor | 2.2E-03 | 2.2E-06 | 1.0E-02 | 5.7E-04 | 1.0E-01 | 5.0E-03 | 0 | 0.55 | — | — |
| Europe | Cyanazine | 3.6E-02 | 3.6E-05 | 1.0E-02 | 5.8E-04 | 2.0E-03 | — | 0 | 0.04 | 0.09 | — | — |
| Endrin | 1.8E-02 | 1.8E-05 | 5.0E-02 | 2.9E-03 | 2.0E-04 | 3.0E-04 | 0 | 0.06 | 0.14 | — | — |
| HCH (gamma) | 7.7E-03 | 7.7E-06 | 2.0E-02 | 1.2E-03 | 8.0E-03 | 3.0E-04 | 0 | 0.05 | 0.16 | 0.01 | — | — |
| Parathion ethyl | 9.2E-03 | 9.2E-06 | 5.0E-02 | 2.9E-03 | 6.0E-04 | 5.0E-03 | 0 | 0.1 | 0.06 | 0 | — | — |
| Phosmet | 4.2E-02 | 4.1E-05 | 1.0E-01 | 5.8E-03 | 1.0E-02 | 4.5E-02 | 0 | 0.05 | 0.15 | 0 | — | — |
| Procymidone | 1.5E-02 | 1.5E-05 | 1.0E-02 | 5.8E-04 | 2.8E-03 | 1.2E-02 | 0 | 0.06 | 0 | 0 | — | — |
| Propazine | 1.4E-02 | 1.4E-05 | 1.0E-02 | 5.8E-04 | 2.0E-02 | 1.7E-02 | 0 | 0.06 | 0.02 | — | — |
| Propyzamide | 1.2E-02 | 1.2E-05 | 1.0E-02 | 5.8E-04 | 2.0E-02 | 7.5E-02 | 0 | 0.07 | 0.01 | 0 | — | — |
| Simetryn | 1.3E-02 | 1.2E-05 | 1.0E-02 | 5.8E-04 | 2.5E-02 | — | 0 | 0.08 | 0 | 0 | — | — |
| Terbutylazine | 2.4E-02 | 2.4E-05 | 1.0E-02 | 5.8E-04 | 4.0E-03 | 8.0E-03 | 0 | 0.08 | 0 | 0 | — | — |
| Terbutryn | 9.5E-03 | 9.5E-06 | 1.0E-02 | 5.8E-04 | 1.0E-01 | 1.0E-03 | 0 | 0.09 | 0.05 | 0.01 | — | — |
| Tetrachlorvinphos | 4.5E-02 | 4.5E-05 | 1.0E-02 | 5.8E-04 | 5.0E-02 | 3.0E-02 | 0 | 0.11 | — | 0.01 | — | — |
| South America | Mirex | 2.6E-05 | 8.3E-09 | 1.0E-02 | 2.5E-04 | — | 2.0E-04 | 0 | 0.53 | 0.54 | — | — |
| Pentachlorobenzene | 8.3E-05 | 2.6E-08 | 1.0E-02 | 2.5E-04 | 1.7E-02 | — | 0 | 0.11 | 0.01 | 0.19 | — | — |
| Continent/pesticide | MEC       | EADI      | MRL      | TMDI     | ADI     | ARfD     | HQ        |
|---------------------|-----------|-----------|----------|----------|---------|----------|-----------|
| **Crustacean**      |           |           |          |          |         |          |           |
| Africa              |           |           |          |          |         |          |           |
| ΣDDD, DDE, DDT     | 2.5E+00  | 1.2E-03   | 1.0E+00  | 2.9E-02  | 5.0E-03 | —        | 0         | 0.04      | 0.24      | —        |
| Chlordane (alpha)   | 3.0E-02  | 1.5E-05   | 2.0E-03  | 5.9E-05  | 5.0E-04 | —        | 0.01      | 0.25      | 0.03      | —        |
| Chlordane (gamma)   | 5.5E-01  | 2.6E-04   | 1.0E-02  | 2.9E-04  | 5.0E-04 | —        | 0.03      | 0.9       | 0.53      | —        |
| Hexachlorobenzene   | 3.6E-02  | 1.7E-05   | 1.0E-02  | 2.9E-04  | 6.0E-04 | 8.0E-04  | 0         | 0.06      | 0.03      | 0.02     |
| Nonachlor (beta)    | 1.9E-02  | 9.2E-06   | 6.0E-03  | 1.8E-04  | —       | —        | 0         | 0.05      | —         | —        |
| **North America**   |           |           |          |          |         |          |           |
| Chlordcone          | 1.4E-01  | 9.9E-05   | 1.0E-02  | 5.8E-04  | —       | —        | 0.01      | 0.17      | —         | —        |
| **Fish**            |           |           |          |          |         |          |           |
| Africa              |           |           |          |          |         |          |           |
| ΣAldrin + dieldrin  | 1.2E+00  | 5.6E-04   | 6.0E-03  | 1.8E-04  | 1.0E-04 | 3.0E-03  | 0.09      | 3.19      | 5.62      | 0.19     |
| ΣDDD, DDE, DDT     | 3.5E+00  | 1.7E-03   | 1.0E+00  | 2.9E-02  | 5.0E-03 | —        | 0         | 0.06      | 0.34      | —        |
| Atrazine            | 6.3E-01  | 3.1E-04   | 1.0E-02  | 2.9E-04  | 2.0E-02 | 1.0E-01  | 0.03      | 1.04      | 0.02      | 0        |
| Carbofuran          | 2.2E-01  | 1.1E-04   | 2.0E-02  | 5.9E-04  | 1.5E-04 | 1.5E-04  | 0.01      | 0.18      | 0.71      | 0.71     |
| Chlordane (alpha)   | 1.8E-01  | 8.7E-05   | 2.0E-03  | 5.9E-05  | 5.0E-04 | —        | 0.04      | 1.48      | 0.17      | —        |
| Chlordane (gamma)   | 1.2E-01  | 5.8E-05   | 1.0E-02  | 2.9E-04  | 5.0E-04 | —        | 0.01      | 0.2       | 0.12      | —        |
| ΣEndosulfan         | 8.6E-01  | 4.2E-04   | 5.0E-02  | 1.5E-03  | 6.0E-03 | 2.0E-02  | 0.01      | 0.28      | 0.07      | 0.02     |
| Endrin              | 4.5E-01  | 2.2E-04   | 5.0E-02  | 1.5E-03  | 2.0E-04 | 3.0E-04  | 0         | 0.15      | 1.09      | 0.73     |
| Endrin aldehyde     | 3.3E+00  | 1.6E-03   | 1.0E-02  | 2.9E-04  | 2.0E-04 | 3.0E-04  | 0.16      | 5.37      | 7.89      | 5.26     |
| HCH (alpha)         | 1.9E+00  | 9.0E-04   | 2.0E-01  | 5.9E-03  | —       | —        | 0         | 0.15      | —         | —        |
| Continent/pesticide | MEC     | EADI    | MRL     | TMDI    | ADI     | ARfD    |
|---------------------|---------|---------|---------|---------|---------|---------|
| HCH (beta)          | 2.3E+00 | 1.1E-03 | 1.0E-01 | 2.9E-03 | —       | —       |
| HCH (gamma)         | 6.6E-01 | 3.2E-04 | 2.0E-02 | 5.9E-04 | 8.0E-03 | 3.0E-04 |
| Heptachlor          | 7.4E-01 | 3.6E-04 | 4.0E-03 | 1.2E-04 | 1.0E-04 | —       |
| Heptachlor epoxide  | 2.5E-01 | 1.2E-04 | 4.0E-03 | 1.2E-04 | 1.0E-04 | —       |
| Hexachlorobenzene   | 5.4E-01 | 2.6E-04 | 1.0E-02 | 2.9E-04 | 6.0E-04 | 8.0E-04 |
| Nonachlor (beta)    | 5.2E-02 | 2.5E-05 | 6.0E-03 | 1.8E-04 | —       | —       |
| Paraquat dichloride | 5.2E+00 | 2.5E-03 | 1.0E-02 | 2.9E-04 | 4.0E-03 | 4.0E-04 |

**Europe**

| Continent/pesticide | MEC     | EADI    | MRL     | TMDI    | ADI     | ARfD    |
|---------------------|---------|---------|---------|---------|---------|---------|
| Aldrin + dieldrin   | 1.0E-02 | 1.0E-05 | 6.0E-03 | 3.5E-04 | 1.0E-04 | 3.0E-03 |
| DDD,DDE,DDT        | 5.5E+00 | 5.5E-03 | 1.0E+00 | 5.8E-02 | 5.0E-03 | —       |
| HCH (gamma)         | 6.0E-02 | 6.0E-05 | 2.0E-02 | 1.2E-03 | 8.0E-03 | 3.0E-04 |
| Hexachlorobenzene   | 5.7E-02 | 5.7E-05 | 1.0E-02 | 5.8E-04 | 6.0E-04 | 8.0E-04 |

**North America**

| Continent/pesticide | MEC     | EADI    | MRL     | TMDI    | ADI     | ARfD    |
|---------------------|---------|---------|---------|---------|---------|---------|
| Chlordane           | 9.6E-03 | 6.9E-06 | 2.0E-03 | 1.2E-04 | 5.0E-04 | —       |
| Chlordane (alpha)   | 1.0E-02 | 7.2E-06 | 2.0E-03 | 1.2E-04 | 5.0E-04 | —       |
| Endrin              | 1.6E-02 | 1.2E-05 | 5.0E-02 | 2.9E-03 | 2.0E-04 | 3.0E-04 |
| Heptachlor          | 7.7E-03 | 5.6E-06 | 4.0E-03 | 2.3E-04 | 1.0E-04 | —       |

**South America**

| Continent/pesticide | MEC     | EADI    | MRL     | TMDI    | ADI     | ARfD    |
|---------------------|---------|---------|---------|---------|---------|---------|
| Aldrin + dieldrin   | 1.5E-01 | 4.7E-05 | 6.0E-03 | 1.5E-04 | 1.0E-04 | 3.0E-03 |

Note: EADI, MRL, TMDI, ADI, ARfD, HQ are abbreviations for reference values for human health in different contexts.
| Continent | Pesticide     | MEC  | EADI  | MRL  | TMDI | ADI  | ARfD | HQ  |
|-----------|---------------|------|-------|------|------|------|------|-----|
|           | Endrin aldehyde | 5.2E-02 | 1.6E-02 | 1.0E-02 | 2.5E-04 | 2.0E-04 | 3.0E-04 | 0.06 | 0.08 | 0.05 |
| Endrin aldehyde | HCH (gamma) | 1.6E-01 | 5.1E-05 | 2.0E-02 | 5.1E-04 | 8.0E-03 | 3.0E-04 | 0.01 | 0.01 | 0.17 |
| HCH (gamma) | Heptachlor     | 3.3E-02 | 1.1E-05 | 4.0E-03 | 1.0E-04 | 1.0E-04 | 0.06 | 0.1 | 0.11 |
| Heptachlor | Heptachlor epoxide | 1.8E-02 | 5.7E-06 | 4.0E-03 | 1.0E-04 | 1.0E-04 | 0.06 | 0.06 |

MEC, measured environmental concentration (mg/kg); EADI, estimated average daily intake (mg/kg bw); MRL, maximum residue limit (mg/kg); TMDI, theoretical maximum daily intake (mg/kg bw/d); ADI, acceptable daily intake (mg/kg bw/day); ARfD, acute reference dose (mg/kg bw/day); body weight (kg): 60.7 (Africa), 57.7 (Asia), 57.7 (Europe), 70.8 (North America), 67.9 (South America); fish and seafood consumption (kg/capita/day): 0.0294 (Africa), 0.05705 (Asia), 0.05765 (Europe), 0.05833 (North America), 0.05745 (South America) and 0.067 (South America).

Table 6. Human health hazard, associated with mollusc, crustaceans and fish consumption, displayed by continent and pesticide.
The highest concentrations were observed in fish from Africa (5.2 mg/kg) and Europe (4.0 mg/kg), followed then by crustaceans in Africa (1.6 mg/kg). The highest number of cases (number of quantifications found considering all the pesticides, species and countries) was registered for fish (169 average cases), followed by molluscs (62 average cases), and finally crustaceans (11 average cases). The elevated number of fish studies is likely due to their importance as a food source.

For allowing a detailed evaluation of human health hazard, the same data is displayed by pesticide and continent and organised considering molluscs, crustaceans and fishes (Table 6). The food consumption rate and the average adult body weight were defined by continent [63, 64]. For the compounds endrin ketone and aldehyde, HCH (sigma and lambda), pretilachlor and pentachlorobenzene, a MRL of 0.01 mg/kg was adopted, since no specific data was found.

Focusing on the molluscs results, the MEC of 15, 52, 10 and 16 pesticides (from Asia, Europe, North America and South America, respectively) were used to calculate the HQs. Due to the low ratio values, only cases with at least one ratio value above 0.05 were presented. As we can see, none of the results proved to be harmful to human through direct consumption. In other words, none of the ratios were above 1, indicating that the calculated EADI was below the reference levels (MRL, TMDI, ADI and ARfD). The highest $\text{HQ}(\text{EADI/TMDI})$ was obtained for methoxychlor in Asia (0.55). For $\text{HQ}(\text{EADI/ADI})$, the highest ratio occurred in South America for mirex with 0.54.

Looking to the crustacean data, a total MEC of eight, two, three and one cases from Africa, Asia, Europe, North America, respectively, were analysed. The same criterion, which is the case with at least one ratio value above 0.05, was applied. High HQs for chlordane (gamma) were observed in crustaceans sampled in Africa (see Table 6). In spite of that, none of the ratios were above 1.

Twenty-four (Africa), 16 (Asia), 10 (Europe), 28 (North America) and 21 (South America) MEC cases were analysed considering the fish data. Once again, only HQ ratios with at least one case above 0.05 are shown. As we can see, none of the maximum average concentrations were above the MRL values; however, several HQ > 1 are observed in Africa, bringing potential exposure for developing carcinogenic health effects. This fact may be a result of bioaccumulation processes (where concentrations increase in higher trophic levels) and/or a higher interest in this matrix (increasing the data availability and diversity). These ratios were registered for six compounds—$\sum$ aldrin + dieldrin, endrin aldehyde, paraquat dichloride, endrin, heptachlor and heptachlor epoxide—where the most preoccupant cases (HQ > 3) are for the first three pesticides cited above.

6. Final considerations

Globally, and because of these high average concentrations, several individual pesticides were quantified at levels exceeding the established LC$_{50}$ for fish and EC$_{50}$ for invertebrates and algae.
In addition, the review has provided clear evidence that the biological data grouped according to Metazoan lineages reached higher concentrations for fish and crustaceans (Figure 3). It is worth noting however that the same pattern was not verified for higher trophic levels including turtles and aquatic mammals which may be due to the lack of samples. Considering that globally, many of the data displayed a wide range of concentrations, coupled with the fact that many of the larger aquatic species are migratory; there is a need to address the pesticide problem from a global perspective.

As a complement to this work, all edible species were evaluated for dietary pesticide risks, as mollusc, crustaceans and fish. No direct human health risk was observed; however, in Africa, some hazard quotients (HQ) were above one, indicating a potential exposure for developing carcinogenic health effects.

In conclusion, the potentially harmful effects of pesticides should be considered not only locally (national/governmental institutions) but also on a global scale.

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