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Less effective selection leads to larger genomes

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Abstract

The evolutionary origin of the striking genome size variations found in eukaryotes remains enigmatic. The effective size of populations, by controlling selection efficacy, is expected to be a key parameter underlying genome size evolution. However, this hypothesis has proved difficult to investigate using empirical datasets. Here, we tested this hypothesis using twenty-two de novo transcriptomes and low-coverage genomes of asellid isopods, which represent eleven independent habitat shifts from surface water to resource-poor groundwater. We show that these habitat shifts are associated with higher transcriptome-wide $d_N/d_S$. After ruling out the role of positive selection and pseudogenization, we show that these transcriptome-wide $d_N/d_S$ increases are the consequence of a reduction in selection efficacy imposed by the smaller effective population size of subterranean species. This reduction is paralleled by an important increase in genome size (25% increase on average), an increase also confirmed in subterranean decapods and mollusks. We also control for an adaptive impact of genome size on life history traits but find no correlation between body size, or growth rate, and genome size. We show instead that the independent increases in genome size measured in subterranean isopods are the direct consequence of increasing invasion rates by repeated elements, which are less efficiently purged out by purifying selection. Contrary to selection efficacy, polymorphism is not correlated to genome size. We propose that recent demographic fluctuations and the difficulty to observe polymorphism variations in polymorphism-poor species can obfuscate the link between effective population size and genome size when polymorphism data is used alone.

Introduction

Eukaryotic organisms exhibit striking variations in their genome size (GS). Within animals, the range of GS extends from 20 Mb in the roundworm Pratylenchus coffeae to 130 Gb in the lungfish Protopterus aethiopicus (Gregory, 2005a). GS shows no correlation with organism complexity, an observation early referred to as the C-value paradox (Thomas, 1971). Even though the contribution of mechanisms such as polyploidization events or transposable element amplification to DNA gain or loss is now better understood (Gregory, 2005b), the evolutionary origin of GS variation still remains largely unexplained (Petrov, 2001).

Large genomes mostly consist of non-coding DNA (Gregory, 2005b; Lynch, 2007). The origins of the large variations in the amount of non-coding DNA found across eukaryotes are currently tentatively explained through two opposite sets of theories. Adaptive theories postulate that variation of the amount of non-coding DNA results in significant phenotypic changes and thus evolves under
the control of natural selection. Main examples of phenotypic changes commonly associated to GS variations include nucleus and cellular sizes (Cavalier-Smith, 1982), growth rate (Grime and Mowforth, 1982) and metabolic rate (Vinogradov, 1995) variations. Conversely, non-adaptive theories postulate that GS variations have little phenotypic impact (Doolittle and Sapienza, 1980), leaving non-adaptive forces such as mutation and genetic drift as the main evolutionary drivers underlying GS variation (Lynch and Conery, 2003). In particular, the mutational-hazard (MH) hypothesis suggests that slightly deleterious mutations, including those that lead to GS variation, can segregate in small populations where the efficacy of purifying selection is impaired by genetic drift (Lynch et al., 2011; Lynch, 2011). Under this hypothesis, the evolution of GS would be controlled by the balance between the emergence of large-scale insertions and deletions (indels) and their fixation rate, which ultimately depends on the efficacy of selection and, thus, the effective population size \( (N_e) \).

Phylogenetic inertia, varying mutational patterns and uncertainties in \( N_e \) estimates, are but a few difficulties that complicate testing of the MH hypothesis with empirical evidence. While GS appears to correlate negatively with population size in eukaryotes (Lynch and Conery, 2003), in agreement with the MH hypothesis, this relationship vanishes when accounting for phylogenetic non-independence among taxa (Whitney and Garland, 2010). In addition, predictions of the MH hypothesis can vary in opposite directions depending on the underlying pattern of indel mutations. In eukaryotes, where indel mutation patterns are typically biased toward insertions, reductions in population size are predicted to lead to increasing GS. Conversely, similar \( N_e \) reductions are expected to result in decreasing GS in bacteria, for which the mutation pattern is biased toward deletions (Kuo et al., 2009). Moreover, although essential to the MH hypothesis, \( N_e \) remains difficult to estimate. Most studies rely on population polymorphism (Lynch and Conery, 2003) or heterozygosity (Yi and Streelman, 2005), two measures that typically reflect population size history over the last tens of thousands to millions of generations, while the pace of genome evolution might take place at much longer temporal scales (Whitney and Garland, 2010; Whitney et al., 2011).

Since the formulation of the MH hypothesis, the few early empirical studies that originally supported a relationship between GS and \( N_e \) (Lynch and Conery, 2003; Yi and Streelman, 2005) have been criticized (Whitney and Garland, 2010; Gregory and Witt, 2008) and later analyses failed to support this relationship. No relationships were found between GS and (i) allozyme polymorphism in plants (Whitney et al., 2010), (ii) molecular polymorphism among different species of rice (genus *Oryza*, Ai et al., 2012), and (iii) the relative population size in seed beetles (Arnaqvist et al., 2015), (iv) genetic drift in salamanders compared to frogs (Mohlhenrich and Mueller, 2016).
Finally, the influence of $N_e$ on *Caenorhabditis* has been dismissed in favor of an adaptive explanation (Fierst et al., 2015). However, all these studies suffered either from the use of very indirect proxies of $N_e$ or from small gene samples, often characterized for not more than 12 species (although exceptions exist, see Whitney et al., 2010). Therefore, the influence of $N_e$ on GS remains to be tested on an empirical dataset that provides a robust estimate of $N_e$ within a statistically powerful framework.

Habitat shifts often result in drastic changes in population size and therefore offer useful case studies for testing the MH hypothesis. In this study, we use a comparative genomic approach to test whether non-adaptive forces drive changes in GS following the habitat shift from surface water to groundwater within asellid isopods. The colonization of groundwater from surface water took place at multiple times and locations over the last tens to hundreds million years ago within this family (Morvan et al., 2013), thereby providing independent replicates of the transition to dark and low-energy habitats (Venarsky et al., 2014; Huntsman et al., 2011). Groundwater colonization leads to eye-degeneration and is considered irreversible (Niemiller et al., 2013). We use 11 pairs of closely-related surface and subterranean asellid species to test the predictions of the MH hypothesis (Table S1). According to the MH hypothesis and assuming that consistent population size reduction took place following groundwater colonization, then, subterranean species are predicted to show reduced selection efficacy and larger GS than their surface relatives. We also considered alternative hypotheses, namely (i) the possible reduction in GS in response to energy limitation in groundwater, and (ii) the selection of particular life history traits (hereafter, growth rate and body size) as a driver of patterns of GS variation.

**Results**

**Efficacy of natural selection in groundwater**

To evaluate differences in selection efficacy between surface and subterranean species, we sequenced and *de novo* assembled the transcriptomes of 11 pairs of asellid species. After gene families delimitation, we estimated the rate of non-synonymous over synonymous substitutions ($d_N/d_S$) on a set of conserved and single copy genes. This ratio is jointly defined by the distribution of selection coefficient of new mutations ($s$) and the magnitude of genetic drift as defined by $N_e$ (Nielsen and Yang, 2003). Therefore, the transcriptome-wide $d_N/d_S$ is expected to increase over extended periods of small $N_e$ because of the increasing fixation of slightly deleterious mutations (Ohta, 1992), an expectation confirmed in a wide range of animals (Galtier, 2016). Consequently, the transcriptome-
wide $d_N/d_S$ is a direct proxy of selection efficacy. Subterranean species show significantly higher transcriptome-wide $d_N/d_S$ than their surface relatives (Figure 1, Table 1). Looking at each pair of species independently, 8 pairs out of 11 display a higher subterranean transcriptome-wide $d_N/d_S$, a relative increase that can be as high as 59% (Figure 1).

While long periods of reduced $N_e$ will induce higher transcriptome-wide $d_N/d_S$, adaptation to new habitats could potentially produce the same effect. Under the action of positive selection, beneficial non-synonymous mutations will reach fixation faster than their synonymous counterparts and will lead to sites with $d_N/d_S > 1$. If the frequency of such sites increases during the transition to groundwater, then, we can expect the transcriptome-wide $d_N/d_S$ to increase. We first tested this adaptive hypothesis using a model that allows $d_N/d_S$ variation across sites and makes it possible to differentiate between variation in the intensity of purifying ($w^-$) and positive selection ($w^+$) and their respective frequencies. Subterranean species do not show an elevated frequency ($f_{q(w^+)}$) or intensity of positive selection ($w^+$) but show higher $w^-$ (Table 1 & S2, Figure S1 & S3). This supports a scenario in which subterranean species do not experience higher rates of positive selection, but instead evolve under reduced purifying selection efficacy.

We next tested the adaptive $d_N/d_S$ increase scenario using polymorphism data. Under high rate of positive selection with recurrent fixation of non-synonymous mutations, populations will display an excess of non-synonymous substitutions compared to non-synonymous polymorphism (McDonald and Kreitman, 1991). We used the “direction of selection” statistics ($DoS$, Stoletzki and Eyre-Walker, 2011), which is a transcriptome-wide comparison of the rates of non-synonymous substitution and polymorphism, to test if positive selection indeed led to higher rates of fixation in subterranean species ($DoS > 0$). Most subterranean species have negative $DoS$ (Figure S4) and subterranean species do not have higher $DoS$ than surface species (Table 1). On the contrary, in most species, irrespective of their habitat, the $DoS$ is close to 0 or negative indicating that many slightly deleterious mutations are segregating in these species. While this observation is in line with the idea that effective population sizes reduces the efficacy of selection in these species, it does not completely rule out the hypothesis that subterranean species may have concomitantly evolved higher $d_N/d_S$ as a result of more frequent adaptations during the shift from surface to subterranean habitats. When slightly deleterious mutations dominate the evolutionary dynamics, which appears to be the case in this group, they can mask the influence of adaptive evolution on polymorphism (James et al., 2016). We further tested this hypothesis by directly estimating the rate of adaptation ($\alpha$) in two species pairs where subterranean species display elevated $d_N/d_S$ than their surface sister species (P. beticus versus P. jalionacus and P. coiffaiti versus P. cavaticus, Figure 1). We used
a McDonald-Kreitman (McDonald and Kreitman, 1991) modified approach (Messer and Petrov, 2013) designed to cancel the influence of demography and linkage effects. For species such as the ones studied here, the high prevalence of segregating deleterious mutations at low frequency inflates the rate non-synonymous polymorphism and artificially decrease \( \alpha \) estimates. Messer and Petrov suggest to reconstruct \( \alpha \) as a function of the derived allele frequencies. As allele frequency increases, slightly deleterious alleles become rarer, to the point that a robust estimate of \( \alpha \) can be obtained by calculating the asymptotic \( \alpha \) for an allele frequency of 1. By re-sequencing the transcriptomes of 4 to 5 individuals per species, we reconstructed the unfolded site frequencies of these two pairs of species, fitted the distribution of \( \alpha(x) \) and estimated the asymptotic \( \alpha(1) \). For each species we recovered the expected distribution of \( \alpha(x) \) as in Messer and Petrov (2013) and obtained asymptotic \( \alpha \) that were negative (Figure S5). While subterranean species of these two pairs of species show clear transcriptome-wide \( d_N/d_S \) increases, they do not display elevated rate of adaptation. For one pair there is no significant \( \alpha \) variation (\( P. beticus - P. jalionacus \), p-value=0.49), while in the other pair the subterranean species shows lower \( \alpha \) estimates (\( P. coiffaiti - P. cavaticus \), p-value=0.02).

Therefore both DoS and \( \alpha \) analyses confirm that the increase in \( d_N/d_S \) in subterranean species is not caused by a higher rate of positive selection, in line with the results found on the model differentiating between variation in the intensity of purifying \((w^-)\) and positive selection \((w^+)\).

Subterranean species share multiple convergent regressive phenotypes, such as the loss of eyes and pigmentation which may ultimately be associated with gene non-functionalizations (Protas et al., 2005; Niemiller et al., 2013). A transcriptome-wide \( d_N/d_S \) increase can therefore also be caused by an excess of genes that have lost their function and consequently have acquired \( d_N/d_S \) nearing 1. As an example, the opsin gene of subterranean species has much higher \( d_N/d_S \) as a result of gene non-functionalization (see next section and Figure S6). Release of functional constraint on a gene and the resulting \( d_N/d_S \) increase should also be paralleled with much lower gene expression, if no expression at all (Zou et al., 2009; Yang et al., 2011). This is typically observed for the opsin gene that has much lower expression in the subterranean species (Figure S6). If gene non-functionalization in subterranean species is pervasive enough to shift the transcriptome-wide \( d_N/d_S \) upwards, we expect to see a subset of genes with lower expression in the subterranean species. We tested this hypothesis by comparing the expression of the genes in the surface and subterranean species of each pair. The set of conserved and single copy genes that was used to calculate each species \( d_N/d_S \) have much higher expression levels than the complete transcriptomes (Figure S7). This set of genes also displays more conserved gene expression levels across species (Figure S7). Finally, after counting changes in gene expression category between sister species, we found no
evidence of an excess of genes with lower expression in the subterranean species (Wilcoxon signed rank test p-value=0.650, Table S4). We further tested this non-functionalization hypothesis by looking for a subset of genes that would display larger $d_{N}/d_{S}$ in the subterranean species while the remaining genes would display no $d_{N}/d_{S}$ variation. Distributions of the variation in gene $d_{N}/d_{S}$ did not support the existence of such a small subset of genes (Figure S2). On the contrary, these distributions were unimodal with a median positively correlated to the transcriptome-wide $d_{N}/d_{S}$ ($R^2 = 0.62$, p-value = 0.004).

Altogether, we accumulated multiple evidences that the transcriptome-wide $d_{N}/d_{S}$ increase observed in subterranean species is not the consequence of increased levels of positive selection or gene non-functionalization, but rather the result of convergent reductions in the efficacy of purifying selection among subterranean species.

**Polymorphism proxies of $N_e$**

Instead of directly assessing selection efficacy, the MH hypothesis has traditionally been tested using polymorphism data. Indeed, polymorphisms provide a direct proxy for $N_e$, which tunes the magnitude of genetic drift and ultimately the efficacy of selection. As transcriptomes were sequenced from pooled individuals, we estimated synonymous and non-synonymous polymorphism for genes with high coverage. We used the population mutation rate ($\hat{\theta}_w$) which is proportional to the product of $N_e$ and the mutation rate $\mu$, and the ratio of non-synonymous over synonymous polymorphism ($p_{N}/p_{S}$), which is expected to decrease with increasing $N_e$, independently of $\mu$. Both the $d_{N}/d_{S}$ and the $p_{N}/p_{S}$ measure the efficacy of selection to purge slightly deleterious mutations, though the later works at a much shorter timescale. As expected, $\hat{\theta}_w$ and $p_{N}/p_{S}$ are negatively correlated (phylogenetic generalized least-squares models, PGLS p-value <0.001, $R^2 = 0.43$). Polymorphism data are generally consistent with selection patterns inferred from $d_{N}/d_{S}$: subterranean species have significantly higher $p_{N}/p_{S}$ than their surface relatives in 6 out of 11 pairs, whereas there is no pair significantly supporting the opposite pattern (Figure 1). However for $\hat{\theta}_w$ the pattern is less clear: in 6 pairs subterranean species have significantly lower $\hat{\theta}_w$, while in 3 pairs subterranean species show significantly higher $\hat{\theta}_w$ (Figure 1). Overall, the differences in $p_{N}/p_{S}$ or $\hat{\theta}_w$ between subterranean and surface species are not statistically significant (Table 1). In addition, there is no correlation between $\hat{\theta}_w$ or $p_{N}/p_{S}$ and the efficacy of selection as estimated using the transcriptome-wide $d_{N}/d_{S}$ (PGLS p-value=0.80 and 0.79, respectively). Traditional polymorphism proxies of $N_e$ do not therefore support the same scenario as the one depicted using selection efficacy ($d_{N}/d_{S}$).
Estimating colonization times with opsin sequences

Subterranean species may have colonized groundwater at different time periods, some being subterranean for a much longer time than others. Ignoring such differences through the use of a qualitative present-day ecological status (i.e. surface versus subterranean) may limit our power to detect a change in GS or polymorphism associated with the subterranean transition. One could contrast polymorphism measures and the time since the latest speciation event, where we know that a species ancestor was a surface species, but this would only be valid if speciation and colonization times were synchronous. Alternatively, we estimated the colonization time using the non-functionalization of the opsin gene. Indeed, similarly to observations made in underground mammals (Emerling and Springer, 2014), together with the regression of the ocular system, some subterranean species display loss-of-function mutations in eye pigment (Leys et al., 2005) or opsin genes (Niemiller et al., 2013), which are indicative of a loss of functional constraint. If we assume that opsin gene sequences have lost their function early in the process of groundwater colonization, then, they must have been evolving under a neutral model \( (d_N/d_S = 1) \) since that colonization. Using a two states model of evolution with one surface opsin \( d_N/d_S \) estimated using opsins from surface species, and one subterranean opsin \( d_N/d_S \) equal to 1, we can then estimate the colonization time as a function of the speciation time and the estimated opsin \( d_N/d_S \) measured on a given branch leading to a subterranean species.

Using a combination of Sanger sequencing, transcriptome assemblies and genome sequencing reads, we reconstructed one opsin ortholog for 19 species out of 22. Irrespective of their ecological status, the two species of the genus Bragasellus probably do not possess this opsin locus. In addition, for one Proasellus subterranean species (P. parvulus), failure to amplify or recover Illumina reads from this locus suggests that the whole locus was lost in this species. Subterranean species showed lower opsin expression levels and had much higher opsin \( d_N/d_S \) ratios than surface species (average \( d_N/d_S = 0.3 \) and 0.05, respectively, Figure S6). In addition to one subterranean species which completely lost the loci (P. parvulus), two subterranean species also harbored clear non-functionalization signatures consisting of an 18 base-long deletion for P. solanasi and the insertion of a 280 base-long repeated element in the sequence of P. cavaticus. These observations validate the opsin locus as a colonization clock.

Estimated colonization times vary a lot with more than 50X variation between the youngest subterranean species (P. jalionacus, 2 MYA) and the oldest one (P. herzegovinensis, 122 MYA, Table S5). Colonization time is related to the regression of the eye and pigmentation, with species with intermediate phenotypes (reduced eyes and partial depigmentation) being very recent sub-
terrestrial species (Figure S8). Using relative colonization time \( \frac{\text{time}_{\text{colonization}}}{\text{time}_{\text{speciation}}} \) for \( d_N/d_S \) ratios or absolute colonization time instead of the present-day ecological status gives very similar results (Table 1), indicating that variations in the colonization time are not likely to obfuscate polymorphism variations. Conversely, the strength of the correlation between the transcriptome-wide \( d_N/d_S \) (or \( w^- \)) and relative colonization time is higher than with the ecological status (Table 1), reinforcing the hypothesis of a causal link between the subterranean colonization and the subsequent drop in selection efficacy. The only exception is the frequency of sites under positive selection \( f_q(w^+) \), Table 1) which becomes significantly higher in subterranean species when colonization time is used instead of ecological status (PGLS p-value=0.016, +0.6% per 100 MY of colonization).

**Genome size increase in groundwater**

We next measured genome sizes in all 11 species pairs using flow cytometry. Using either the ecological status or the colonization time, we found a statistically significant increase in GS following the transition from surface to groundwater habitats (Figure 1, Table 1). Looking at each pair of species independently, 7 pairs out of 11 display a significantly higher GS, a relative increase reaching 137% in *P. hercegovinensis* (Figure 1). This finding is robust to the addition of 19 asellid species (PGLS with 41 asellids, p-value= 0.022, coefficient=0.273) and to the inclusion of a wider range of metazoans (linear mixed model with 18 independent pairs including Decapoda and Gastropoda, p-value = 0.040, 25% average increase in GS, Figure 2, Table S6).

**Linking genome size to selection efficacy**

One of the main predictions of the MH hypothesis is that GS is negatively correlated with selection efficacy in eukaryotes. We validated this prediction because we found a highly significant positive relationship between GS and the transcriptome-wide \( d_N/d_S \) (or \( w^- \), Table 1, Figure S9). In addition the \( d_N/d_S \) ratio (or \( w^- \)) achieves similar, if not better, performance in predicting GS variation than the ecological status or colonization time (lower AIC and higher R\(^2\), Table 1). When \( d_N/d_S \) (or \( w^- \)) is put first and ecological status second into a single PGLS model of GS, the effect of the ecological status is no longer significant (PGLS p-value=0.189).

**Testing other covariates**

In contradiction to the MH hypothesis, adaptative hypotheses postulate that variation in GS is under direct selection via its impacts on cellular (such as nucleus and cell sizes) and organismal
parameters (such as body size and growth rate) (Gregory, 2001). In many species, population size
covaries with traits under selection such as growth rate and body size, themselves correlated to some
extent to GS, making any causation test extremely challenging (Gregory, 2005b). While body size
was readily available in the literature, we estimated growth rate in 16 species using the RNA/protein
ratio which is known to be positively correlated to growth rate in Rotifera (Wojewodzic et al., 2011).
Indeed, in situ estimates of growth rates were out of reach, and a more traditional proxy such as
the RNA/DNA ratio is inapplicable when GS varies. In accordance to the general assumption that
subterranean animals tend to adopt K-selection life history traits, subterranean asellids species
display lower growth rate, though no trend was found regarding body size (Table 1). However,
growth rate and body size do not correlate with GS (Table 1, Figure S9). Thus, although many
forces might be at play during the transition to groundwater habitats, in asellids, we only found
correlation between selection efficacy and GS.

Mechanism of genome size increase

Implicit in the MH hypothesis is that an increase in GS should result from the progressive spread
of insertions with slightly deleterious fitness effects, such as transposable elements (Vieira et al.,
2002). Yet, other much faster mechanisms such as polyploidization events can also inflate GS (Otto,
2007). We tested for the occurrence of such large duplication events by looking for an excess of
recent paralogs in the 11 subterranean species compared to their surface sister species. The mean
number of gene copy per gene family is not correlated to the ecological status nor to GS (PGLS
p-value=0.773 and 0.579, respectively, Figure S10), indicating that subterranean species do not
present an excess of recent duplication events.

To test for the accumulation of repeated elements, we evaluated the amount of repeated DNA
in the 11 asellid species pairs using low-coverage genome sequencing, followed by clustering of the
highly repeated elements. Indeed, contrary to the non-repeated fraction of the genome, elements at
high frequency will collect enough reads to be assembled. Summing across the contributions of each
element provides an estimate for the size of the repeatome (ie. the fraction of the genome consisting
of repeated DNA elements). We found larger repeatomes in large genomes (Figure 3A & D, Table 2,
Figure S11). The repeatomes are largely made of repeat families found in a single species, called
repeat orphans, with very few shared repeats across species (Figure 3B). The occurrence of these
shared repeats is largely explained by phylogenetic relatedness: closely related sister species share
more than 200 repeat families with this number quickly decreasing with divergence time (Figure 3C).
GS has little power to explain the composition of the repeatome. None of the axes of a repeatome
composition correspondence analysis is correlated to GS while the first three axes harbor a strong phylogenetic signal (Blomberg $K > 1$ with $p$-value < 0.01, Table S7, Figure S12).

The pattern of GS increase is globally congruent with a global increase of the repeatome invasiveness. Indeed, big genomes have at the same time more repeats and repeats at higher frequencies (Table 2, Figure 3E & D). To a lesser extent, the number of repeat orphans and their frequencies also increase with GS (Table 2), demonstrating that big genomes are also more prone to genome invasion by new repeats. Nonetheless, the ratio of the total genomic size (TGS) occupied by new repeats over common repeats do not change ($\frac{TGS_{orphans}}{TGS_{non-orphan}}$, Table 2), indicating that this aspect of the repeat community structure does not change as GS increases. So, contrary to several model organisms such as humans or maize, the GS increase was not induced by a very limited set of elements, but is the consequence of a repeat element community that became globally more invasive subsequently to the ecological transition.

While the repeated portion of the genome increases linearly with GS, it does not explain 100% of GS variations: on average 1Gb of repeats was gained for 1.3Gb of GS increase (Table 2). Consequently, the estimated TGS of the non-repeated portion of the genome also increases with GS, though at a much slower pace (1Gb for 2.8Gb, Table 2). Either repeats are harder to assemble in large genomes or another minor mechanism is also at play during GS increase. Directly using the repeatome size instead of GS in correlation analyses gives similar or reinforced results: while polymorphism based $N_e$ proxies ($\hat{\theta}_w$ or $p_N/p_S$), growth rate and body size do not correlate with repeatome size, selection efficacy ($d_N/d_S$ or $w^-$) does (Table 1).

**Discussion**

We found a substantial correlation between selection efficacy, as measured by transcriptome-wide $d_N/d_S$, and repeatome size. This finding indicates that, for a large part, GS is controlled by the efficacy of selection to prevent the invasion of the genome by repeated elements. Conversely, we found no correlation between $N_e$ estimates derived from polymorphism data and GS. At first glance, this result sounds contradictory since the efficacy of selection depends on $N_e$. We propose two non-mutually exclusive hypotheses to explain this contradiction. First, while the transcriptome-wide $d_N/d_S$ provides an average estimate of selection efficacy since the divergence of two species of a pair, polymorphism-based proxies such as $\hat{\theta}_w$ or $p_N/p_S$ are influenced by recent $N_e$ fluctuations, independently of the divergence time. If $N_e$ fluctuates rapidly with large amplitude around a stable mean, polymorphism is likely to provide a noisy proxy of this mean, contrary to the $d_N/d_S$. 
This hypothesis is supported by larger coefficients of variations for $\hat{\theta}_w$ or $p_N/p_S$ than for the $d_N/d_S$ ($CV_{d_N/d_S} = 0.22$, $CV_{p_N/p_S} = 0.39$, $CV_{\hat{\theta}_w} = 0.73$, test of the equality of CVs p-value < 0.001). In this study, we observed the effect of multiple groundwater transitions that happened tens to hundreds million years ago, a time scale long enough to produce large $d_N/d_S$ and GS variations, but also encompassing important climatic fluctuations which potentially generated shorter time-scale $N_e$ variations. Particularly, quaternary climatic fluctuations are likely to have produced important $N_e$ fluctuations in surface species which have much more unstable habitats than subterranean species. Therefore, the lack of a clear correlation between short-term $N_e$ proxies, like polymorphism, and GS might be the consequence of recent climatic fluctuations. Interestingly, this hypothesis received some support using Tajima’s D tests. Indeed, only $P. beticus$ (a surface species) out of the 4 species for which we have adequate data to estimate Tajima’s D is not at the mutation-drift equilibrium (p-value=0.017, Table S8). This surface species shows evidence of recent population contraction (Tajima’s D > 0) which might explain its unexpected combination of low polymorphism and low $d_N/d_S$ when compared to its sister subterranean species (Figure 1). Second, polymorphism variation might also be more prone to measurement artifacts than $d_N/d_S$. In particular, SNP calling errors can constitute a relatively important fraction of detected variants among species with low levels of polymorphism. For most species pairs (7 out of 11), we observed a higher level of polymorphism in surface than in subterranean species (Figure 1, Figure S13). The four other cases all correspond to pairs for which both species have low level of polymorphism, which might therefore be subject to higher measurement error rates. This is in line with the strong relationship observed between the surface species $\hat{\theta}_w$ and the difference in $\hat{\theta}_w$ between species pairs ($R^2 = 0.74$, p-value < 0.001, Figure S13). This suggests that below approximately 2% $\hat{\theta}_w$ becomes a poor indicator of $N_e$ changes. Altogether, the use of polymorphism proxies of $N_e$ for polymorphism-poor species or for species which experienced recent $N_e$ variations might therefore result in misleading rejection of the MH hypothesis.

Our findings shed new light on the debate of the validity of the MH hypothesis and the comparative methods that should be implemented to test it (Charlesworth and Barton, 2004; Gregory and Witt, 2008; Whitney and Garland, 2010; Lynch, 2011; Whitney et al., 2011). Using a relatively reduced set of ecologically-contrasted species pairs as true replicates of the same ecological transition is statistically more powerful than testing for differences in genomic attributes among a larger set of distantly related taxa, in which the number of independent observations is unknown and for which many traits varies. Accounting for phylogenetic effects in statistical analyses of GS variation among multiple species is another yet crucial aspect because it increases not only specificity (Whit-
ney and Garland, 2010) but also sensitivity. Taken altogether, subterranean species do not have larger GS than surface species (ordinary least square linear model p-value = 0.122 for the 11 species pairs, p-value = 0.095 for the 41 asellid species, p-value = 0.261 for the 18 metazoan species pairs), whereas pairwise comparisons of surface and subterranean species (Figure 1 & 2) and PGLS models (Table 1) reveal a very clear and significant pattern of higher GS among subterranean species.

While there is several lines of evidence supporting a lower selection efficacy caused by long-term \( N_e \) reduction in subterranean species, we found little support that the colonization of this new habitat was also paralleled with adaptive evolution. The only evidence was found in the frequency of sites under positive selection when colonization time was used. The increase was nonetheless moderate (+0.6% per 100 MY of colonization) and was not supported by polymorphism (DoS or \( \alpha \)) analyses. However, this study is limited to a small set of gene families that are found in most species, in a single copy and whose expression is very conserved. This set of genes is probably under strong purifying selection and might be less prone to positive selection. Fully investigating the relative role of adaptive versus non-adaptive forces during this ecological transition will require a much broader genomic approach.

Disentangling the forces that drive GS variation has commonly been complicated by rampant covariation between GS and multiple traits such as cell and body sizes, growth rates, metabolism and \( N_e \), to name a few. In this study, we found no association between GS and two common covariates (body size and growth rate). While we cannot completely rule out other non-tested parameters and alternative ad hoc adaptive hypotheses, the results of this study are fully compatible and best explained by a causal relationship between \( N_e \) and GS. The mechanisms that drive genome size variation are also fully compatible with the MH hypotheses. The repeated elements were globally more diversified and more invasive in species with reduced selection efficacy, an expected outcome if selection against repeated element proliferation is less effective.

Documenting changes in the architecture of genomes among taxa that have undergone major shifts in habitats (Protas et al., 2005; Jones et al., 2012; Fang et al., 2014; Soria-Carrasco et al., 2014) holds much promise for disentangling evolutionary processes driving genome evolution. According with the MH hypothesis, our focus on the genomics of groundwater colonization brings new evidence for a prominent role of non-adaptive forces in GS evolution. Despite strong energetic constraints in groundwater, GS likely increases under the long-term effect of reduced \( N_e \), which limits the strength of natural selection in hampering the invasion of slightly deleterious repeated elements. Altogether this study supports long-term effective population size variation as a key evolutionary regulator of genome features.
Methods

Aselloidea timetree

Phylogenetic comparative methods require accurate estimates of phylogenetic relationships and divergence times among species (Purvis et al., 1994). Both were inferred from a large timetree of Aselloidea containing 193 evolutionary units (Morvan et al., 2013) (Table S9). Sequences of the mitochondrial cytochrome oxidase subunit I (COI) gene, the 16S mitochondrial rDNA gene and the 28S nuclear rDNA gene used to build the Aselloidea timetree were obtained according to methods described in (Calvignac et al., 2011; Morvan et al., 2013). Alignments and Bayesian estimates of divergence times were conducted according to (Morvan et al., 2013). From the Aselloidea timetree, we selected 11 independent pairs of surface and subterranean asellid species as replicates of the ecological transition from surface water to groundwater.

RNA-seq

For the 11 selected species pairs (Table S1) individuals were sampled from caves, springs, wells, and the hyporheic zone of streams using different pumping and filtering devices (Bou-Rouch pump, Cvetkov net, and Surber sampler) and were flash frozen alive. Total RNA was isolated using TRI Reagent (Molecular Research Center, Cincinnati, USA). Extraction quality was checked on a Bioanalyser RNA chip (Agilent Technologies, Santa Clara, USA) and RNA concentrations were estimated using a Qubit fluorometer (Thermo Fisher Scientific, Waltham, USA). Prior to any additional analysis, species identification was corroborated for each individual by sequencing a fragment of 16S gene. Equimolar pools of at least 5 individuals were made to achieve 10µg of the total RNA (Table S1). Volumes were reduced using a Concentrator-Plus (Eppendorf, Hamburg, Germany) to achieve approximately 10µL. Double strand poly(A)-enriched cDNA were then produced using the Mint2 kit (Evrogen, Moscow, Russia) following the manufacturer protocol except for the first-strand cDNA synthesis, where the CDS-1 adapter was used with the plugOligo-Adapter of the Mint1 kit (5’-AAGCAGTGGTATCAACGCAGAGTACGGGGG_P_3’). After sonication with a Bioruptor Nextgen UCD300 (Diagenode, Lige, Belgium) and purification with MinElute (Qiagen, Hilden, Germany), Illumina libraries were prepared using the NEBNext kit (New England BioLabs, Ipswich, USA) and amplified using 22 unique indexed primers. After purification with MinElute, 400-500 base pair fragments were size selected on an agarose gel. Libraries were paired-end sequenced on a HiSeq 2000 sequencer (Illumina, San Diego, USA) using 100 cycles at the Danish National High-throughput DNA Sequencing Center (Copenhagen, Danemark). A full lane was used for one species (Proasellus beticus), and
reads were resampled to represent about 2%, 5%, 10%, 25%, 50% and 100% of a full lane. These 6 sets of reads were de novo assembled (see next section) and the number of assembled components > 1kb was compared among sets (Figure S14). This preliminary experiment was used as a rational procedure to multiplex 4 species on one lane.

**Transcriptome assembly**

Adapters were clipped from the sequence, low quality read ends were trimmed (phred score <30) and low quality reads were discarded (mean phred score <25 or if remaining length <19bp) using fastq-mcf of the ea-utils package (Aronesty, 2013). Transcriptomes were de novo assembled using Trinity (version 2013-02-25, Grabherr et al., 2011). Open reading frames (ORF) were identified with TransDecoder (http://transdecoder.sourceforge.net/). For each assembled component, only the longest ORF was retained, and gene families were delimited using all against all BLASTP (Altschul et al., 1990) and SiLiX (Miele et al., 2011). SiLiX parameters were set to i=0.6 and r=0.6 as they were maximizing the number of 1-to-1 orthologous gene families.

**d\textsubscript{N}/d\textsubscript{S} calculation**

Single copy orthologs were extracted for 3 different sets of taxa: the 11 asellid species pairs (320 genes), the ibero-aquitanian clade (863 genes), and the alpine-coxalis clade (2257 genes) (see Figure 1). Each gene family was then aligned with the following procedure: (1) search and masking of frameshift using MACSE with frameshift cost set to -10 (Ranwez et al., 2011), (2) multiple alignments of the translated sequences with PRANK (Löytynoja and Goldman, 2008) using the empirical codon model and F option, (3) site masking with Gblocks (Castresana, 2000) using -t=c, -b5=h and -b2 set as -b1. After gene concatenation, a transcriptome-wide d\textsubscript{N}/d\textsubscript{S} ratio was estimated with the free ratio model of CODEML from the package PAML 4.7a (Yang, 2007). Confidence intervals were obtained using 100 non-parametric bootstrap samples (random sampling of the codon sites with replacement).

To test whether the observed d\textsubscript{N}/d\textsubscript{S} increase could be attributed to a reduction in the efficacy of purifying selection or to an increase in positive selection (a higher number of sites under positive selection and/or an elevated positive selection intensity), we used the M10 branch-site model (Yang et al., 2000) as implemented in BppML (Dutheil and Boussau, 2008). To reduce computation times, the analysis was performed using quartets of taxa: the two species of a pair and two additional surface species used to root the tree. Large alignments (>300 000 codons) were reduced by randomly sampling 280 000 codons and only sites that were complete were retained. Purifying and positive
selection intensity ($w^-$ and $w^+$) and frequency ($fq(w^-)$ and $1 - fq(w^-)$) were estimated using the a posteriori mean site $d_N/d_S$ using BppMixedLikelihoods (Dutheil and Boussau, 2008).

**Single nucleotide polymorphism**

Estimating population polymorphism from pooled RNA-seq samples is complicated by the fact that (1) RNA-seq is prone to both RT-PCR and sequencing errors (Gout et al., 2013), (2) polymorphism can be over-estimated by hidden paralogs (Gayral et al., 2013), and (3) it is difficult to differentiate low frequency alleles from sequencing errors in pooled data-sets (Futschik and Schlötterer, 2010).

While an accurate estimate is currently out of reach, it is possible to obtain polymorphism estimates that are comparable across taxa. We developed a statistical design that (1) is conservative in defining polymorphism, (2) balances the sampling effort so that estimates obtained within a species pair are comparable and (3) maximizes the number of analyzed genes to gain statistical power. Single nucleotide polymorphism (SNP) was searched on a set of 5027 gene families that were present as a single copy in at least 6 out of the 11 species pairs. Gene families with hidden paralogs were filtered by using 10X coverage DNA-seq data available for 4 species (P. karamani, P. hercegovinenis, P. ibericus and P. arthrodilus). Gene families that had a DNA-seq coverage higher than the 90th percentile in any of these 4 species were filtered for any further polymorphism analysis. RNA-seq reads were aligned on the assembled ORF using BWA (aln algorithm, (Li and Durbin, 2009)). SAMtools (Li et al., 2009) was then used to generate a BAM file, discard duplicated reads, and export a BCF file. SNPs were filtered and called with bcftools and vcfutils.pl with the following conservative filtering parameters: minimum read depth of 10, minimum number of reads supporting an allele of 4, and minimum distance to a gap set to 15. Then, SNPs were classified as synonymous or non-synonymous. Only the genes that were highly covered in both species of a pair (average coverage >50X, same results were obtained with lower coverage cutoff) were further considered to compute transcriptome-wide summary statistics (Table S3). This design ensured that synonymous and non-synonymous polymorphism estimates could be compared across taxa, although each of these estimates might be over or under estimated. We then calculated the population mutation rate ($\theta$) using the Watterson estimator:

$$\hat{\theta}_w = \frac{p_S}{\sum_{i=1}^{2n-1} \frac{1}{i}}$$

with $p_S$ the frequency of synonymous segregating sites and $n$ the number of pooled individuals. Finally, we calculated the ratio of non-synonymous over synonymous segregating sites $p_N/p_S$. Con-
Confidence intervals were obtained by bootstrapping the genes 10000 times.

This polymorphism data-set was also used to measure the direction of selection statistics (DoS, Stoletzki and Eyre-Walker, 2011) using:

$$\text{DoS} = \frac{D_n}{D_n + D_s} - \frac{P_n}{P_n + P_s}$$

with $D_s$ and $D_n$ the number of synonymous and non-synonymous divergences, and $P_s$ and $P_n$ the number of synonymous and non-synonymous polymorphisms. Divergences were measured with PAML 4.7a (Yang, 2007) and polymorphisms using the above described pipeline. DoS were measured gene by gene for every species pairs and compared for every pair using a Wilcoxon signed rank test or globally using the median DoS per species.

**Rate of adaptation and Tajima’s D**

For two species pairs ($P. beticus$, $P. jalionacus$, $P. coiffaiti$ and $P. cavaticus$), we performed additional RNA-seq 50 base single-end Illumina sequencing, but this time by independently sequencing 4 to 5 individuals per species owing the estimation of allele frequencies. Reads were mapped on the assembled transcriptomes using BWA (mem algorithm, (Li and Durbin, 2009)), and SNPs were called using Reads2snp (Gayral et al., 2013). The site frequency spectrum (SFS) were then unfolded using the alignment with the respective sister species orthologs. Only sites with non-ambiguously reconstructed ancestral and derived allele were kept. We then used Messer and Petrov approach (Messer and Petrov, 2013) to directly estimate the proportion of adaptive substitutions ($\alpha$) from the unfolded-SFS. Confidence intervals for $\alpha$ were obtained by bootstrapping the SNP 1000 times.

The same dataset was also used to test if populations were at the mutation-drift equilibrium using Tajima’s D test (Tajima, 1989).

**Colonization time**

For 19 species, we were able to determine the sequence of one opsin gene. Sequences were determined using (i) transcriptome sequences, (ii) Sanger sequencing using (Taylor et al., 2005) PCR primers (LWF1a and Scylla) and PCR conditions, and (iii) genomic Illumina reads as detailed below. For the latter, reads were mapped on the closest available opsin sequence following the same approach as for the SNP search and a consensus was called with the Samtools program suite.

To estimate colonization time, we used the loss of function observed in several subterranean species and postulated that the opsin gene loss-of-function took place at the time of groundwater
colonization. We used a model with two $d_N/d_S$ ratios, one for the functional opsins ($\omega_{surf}$) and one for the non-functional opsin ($\omega_1$) which was set to 1. We then defined the $d_N/d_S$ of a branch leading to a subterranean taxa ($\omega_{subt}$) as the weighted mean between these two defined ratios, such as:

$$\omega_{subt} = \omega_{surf} \frac{T - t}{T} + \omega_1 \frac{t}{T}$$

with $T$ the speciation time, and $t$ the time of colonization. From this, we estimated the time of colonization as follows:

$$t = T \times \frac{\omega_{subt} - \omega_{surf}}{1 - \omega_{surf}}$$

Another relevant parameter is the proportion of time a species has been subterranean since the divergence with its sister species, named the relative colonization time ($RCT$), which we estimated using:

$$RCT = \frac{t}{T} = \frac{\omega_{subt} - \omega_{surf}}{1 - \omega_{surf}}$$

Opsin $d_N/d_S$ were estimated using PAML free-ratio branch model, and the $\omega_{surf}$ was estimated as the average $d_N/d_S$ of the surface species showing the most obvious surface phenotypes (P. coiffaiti, P. coxalis, P. karamani, P. ibericus, P. meridianus and P. beticus).

Measurement of genome size

We measured GS for 41 species of asellid (including the 11 species pairs), 4 Atyidae (Pancrustacea, Decapoda) and 2 Rissoidae (Mollusca, Gastropoda) (Table S6). After sampling, individuals were preserved at ambient temperature in silica gel. Measurements were conducted according to (Vieira et al., 2002). Nuclei were extracted from the head of organisms (or from entire individuals when body size was less than 3 mm in length). Heads were crushed in 200µL of cold modified Galbraiths nuclei isolation buffer (20 mM MOPS, 20.5 mM MgCl2, 35.5 mM trisodium citrate, 0.1% Triton X-100, 20µg mL$^{-1}$ boiled RNase A, pH 7.2 adjusted with NaOH; (Galbraith et al., 1983)). The mixture was filtered through 100µm and then 30µm mesh-size nets. The filtrate was centrifuged during 10 sec at 2600g and the supernatant was carefully removed. Pellets were resuspended in 200µL of nuclei isolation buffer. The resuspension was again centrifuged during 10 sec at 2600g and the supernatant was carefully removed. Pellets were resuspended in 250µL of buffer and transferred in 5 mL polystyrene round-bottom tubes. An amount of 50µL of propidium iodure was added to each tube. Tubes were kept in ice and darkness until GS measurements.

Genome sizes were measured using FACSCanto II flow cytometer (Becton Dickinson Instru-
ments) fitted with an argon laser at 488-nm wavelength. We analyzed 5 individuals per species. Individuals were measured in a random order and two individuals of the same species were never analyzed in the same run. Samples were calibrated to two external standards: *Drosophila virilis* females (GS of 0.41 pg, Bosco et al., 2007) and *Asellus aquaticus* (GS of 2.49 pg, Rocchi et al., 1988, and authors’ cross validation). The *Drosophila* were maintained under laboratory conditions at 25 °C for two to three generations before GS measurements. Standards were prepared using the protocol described above from 5 organism heads and were measured in each run (2 measures for *D. virilis* at the beginning and end of runs and 5 measures for *A. aquaticus* evenly distributed during the runs).

The FlowQ bioconductor package (Gentleman et al., 2004) in R software (R Core Team, 2013) was used for quality assessment of flow cytometry data. All the flow cytometer analyses were checked for cell number, boundary events and time anomalies. Cell subsetting known as gating, was firstly performed manually using the BD FACSDiva software (BD Biosciences). Secondly, the automatic curvHDR filtering method (Naumann et al., 2010) was used to select the cells located in the highest density region (HDR level = 0.8). Then, when drift over time was significant, gated values were corrected using a linear regression on *A. aquaticus* reference using the following equation: $IP_{corr} = IP_x - \frac{IP_{st}}{x} \times (x - x_0) \times \lambda$, where $IP_{corr}$ is the corrected gated value, $IP_x$ is the gated value to correct, $IP_{st}$ is the *A. aquaticus* reference gated value at the beginning of the run (time $t = x_0$) estimated by the linear regression, $x$ is the measurement time for the gated value to correct, $x_0$ is the time at the beginning of the run and $\lambda$ is the slope of the linear regression on *A. aquaticus* reference. Drift was considered significant when the regression on *A. aquaticus* reference had adjusted R$^2$ values $\geq 0.1$. Finally, GS were derived from fluorescence data using *D. virilis* as a standard for Asellidae and Rissoidae and using *A. aquaticus* as a standard for Atyidae. Indeed, large GS in Atyidae (previously known GS range from 3.30 to 7.20 pg, Gregory, 2005a) prevented the use of *D. virilis* as a standard.

**Growth rate and body size**

Growth rates were estimated using the total RNA normalized by the total protein biomass of an organism (RNA/protein ratio, Wojewodzic et al., 2011) for at least 7 individuals per species. Total RNA and proteins were isolated using TRI Reagent (Molecular Research Center). RNA concentrations were calculated by fluorometry with a Qubit (Life Technologies). Total proteins were obtained using the Bicinchoninic acid assay (Smith et al., 1985). Body size were estimated using maximum body size as reported in each species description.
To compare the size of the repeatome between surface and subterranean species, we sequenced the genome of the 11 pairs of Asellidae species. For 4 species (*P. ibericus, P. arthrodilus, P. karamani, P. hercegovinensis*), we built blunt-ended libraries for shotgun sequencing on Illumina platforms, as described in (Orlando et al., 2013; Seguin-Orlando et al., 2013) with few modifications. One µg of DNA in 100µL TE buffer was sheared using a Bioruptor NGS device (Diagenode) with four cycles of 15 seconds ON/90 seconds OFF. The obtained size distributions of sheared DNA fragments were centered at around 500 bp. After concentration in 22µL EB buffer (Qiagen) with the MinElute PCR Purification kit (QIAGEN), the sheared DNA fragments were built into blunt-ended DNA libraries using the NEBNext Quick DNA Library Prep Master Mix Set for 454 (New England BioLabs, reference nb. E6070L), following the protocol described in (Meyer and Kircher, 2010), but with 0.5µM Illumina adapters (final concentration). All reactions were carried out in 25µL volumes; incubation times and temperatures were as follows: 20 min at 12°C, 15 min at 37°C for end-repair; 20 min at 20°C for ligation; 20 min at 37°C, 20 min at 80°C for fill-in. After the end-repair and ligation steps, reaction mixes were purified using the MinElute PCR Purification Kit (Qiagen) using elution volumes of 16µL and 22µL of EB buffer, respectively. The final 25µL volume of blunt-end libraries was split in two parts and PCR amplified independently in 50µL reaction mixes containing: 5 units Taq Gold (Life Technologies), 1X Gold Buffer, 4 mM MgCl2, 1 mg/ml BSA, 0.25 mM of each dNTP, 0.5µM of Primer inPE1.0 (5’-AAT GAT ACG GCG ACC ACC GAG ATC TAC ACT CTT TCC CTA GAC GAC GCT CTT CCG ATC T-3’) and 0.5µM of an Illumina 6 bp-indexed (I) primer (5-CGA GCA GAC GAC ATA GGA GAT III III CTO ACT GGA GTC TAC CCT CCG G). Thermocycling conditions for the amplifications were: activation at 92°C for 10 min; followed by nine cycles of denaturation at 92°C for 30 sec, annealing at 60°C for 30 sec, elongation at 72°C for 30 sec; and final elongation at 72°C for 7 min. PCR products were purified using the MinElute PCR Purification kit, with a final elution volume of 25µL EB buffer.

For the remaining 18 species, we built Illumina TruSeq DNA PCR-free LT libraries (Illumina, catalog FC-121-3001), following manufacturers recommendations. Briefly, 1µg of DNA extract was sheared in a total volume of 50µL TE buffer, using a Bioruptor NGS device (Diagenode) with three cycles of 25 seconds ON/90 seconds OFF. The fragmented DNA was cleaned up using Illumina Sample Purification Beads. After end-repair, the DNA fragments were size-selected around 350bp using two consecutive bead purification steps, A-tailed, and ligated to 6bp-indexed Illumina TruSeq adaptors (Set A). Two last bead purifications were performed to remove any adaptor dimer and the final libraries were resuspended in a volume of 20µL Resuspension Buffer. In order to control
for contaminations, library and PCR blanks were carried out at the same time as the samples. Amplified libraries and blanks were quantified using the 2100 Bioanalyzer (Agilent) High-Sensitivity DNA Assay. No detectable amount of DNA could be recovered from the blanks. Blunt-End indexed DNA libraries were pooled and sequenced on two lanes of a HiSeq 2000 Illumina platform (100 cycles paired-end mode run), while the two PCR-free libraries pools were sequenced each on one flow-cell of the Hiseq 2500 Illumina platform (150 paired-end run, Rapid Mode, 6bp index read), at the Danish National High-Throughput DNA Sequencing Centre.

**Repeatome size estimates**

We used low coverage read sequencing to characterize repeated genome sequences (Novák et al., 2010) using RepeatExplorer (Novák et al., 2013). Prior to analysis, DNA-seq reads were randomly sampled to achieve 0.05X coverage following the GS estimated by flow cytometry, so that estimates are comparable across taxa. After clustering of the reads into highly repeated elements by RepeatExplorer, the number of reads representing each repeated element is a direct function of the repeat frequency, the GS, and the sequencing effort. The proportion of the genome (GP) made by this repeat is \( GP_i = n_{reads_i} / n_{reads} \), with \( n_{reads_i} \) the number of reads making the repeat \( i \), and \( n_{reads} \) the total number of reads. The proportion of genome made of repeats (the repeatome size) is then \( GP = \sum_i GP_i \). By default, RepeatExplorer filters repeat elements that have genome proportion lower than 0.01%. To achieve comparable estimates independent of genome sizes, the number of repeats or the repeatome size of a given genome were recalculated by filtering repeats that occupied less than 0.5 Mb. A repeat element total genomic size (TGS) was then calculated as \( TGS_i = GP_i \times GS_j \) with \( GS_j \) the genome size of species \( j \). Repeat families were delimited using blastn (e-value=0.1) and SiLiX (i=50 and r=70).

**Comparative analyses**

Phylogenetic generalized least squares (PGLS) regression models (Martins and Hansen, 1997) were used to test for the correlation between 2 variables. We first tested the association between the ecological transition and population size, biological traits or GS (top of Table 1). PGLS were also used to test for the association between GS and population size, biological traits (bottom of Table 1) or genomic features (Table 2). The correlation between two variables was assessed by comparing a model without the predictor variable (intercept model) to a model including the predictor variable using a likelihood ratio test (LRT). Analyses were performed in R using ape (Paradis et al., 2004) and nlme (Pinheiro et al., 2014) packages. The best model of trait evolution and its associated
covariance structure, in our case the Brownian motion model, was selected according to minimum Akaike information criterion (AIC). The difference in $\theta_w$, $p_N/p_S$, and $d_N/d_S$ between the two species of a surface - subterranean pair was tested using the proportion of bootstrap replicates supporting a difference (critical level = 5%) and the difference in GS was tested using a Wilcoxon rank sum test with 5 measurements of GS (ie. 5 individuals) per species. We also performed ordinary least squares models to test for the effect of the ecological status on GS while ignoring phylogenetic relationships among species. To test for the effect of the ecological transition on GS using a wider range of taxa (ie. 18 species pairs including Decapoda and Mollusca with 5 measurements of GS per species), we performed a linear mixed model in R using the nlme package because a chronogram with accurate branch length could not be obtained given the available molecular markers and calibration points. The ecological status (ie. surface versus subterranean) was a fixed effect, and we specified the random error structure as ecological status nested into species pairs to account for phylogenetic relationships among species. Then, we performed the model with no hierarchy in the random error structure, which is equivalent to an ordinary least squares model, to test for the effect of the ecological status on GS while ignoring phylogenetic relationships among species. Differences in the coefficient of variations of different variables were tested using modified signed-likelihood ratio test (Krishnamoorthy and Lee, 2014) using the R package cvequality.

Data access

Sequence reads and assemblies have been deposited to the European Nucleotide Archive (ENA; http://www.ebi.ac.uk/ena) under the study accession number PRJEB14193. Sanger sequences were submitted to NCBI (GenBank; https://www.ncbi.nlm.nih.gov/nucleotide/) under the accession numbers KC610091-KC610505 (Table S9).

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Disclosure declaration

The authors declare no conflicts of interest.

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Figures

Figure 1:
| Genome size (pg) | 1.0 | 1.5 | 2.0 | 2.5 | 3.0 | 3.5 |
|------------------|-----|-----|-----|-----|-----|-----|
| 70%              |     |     |     |     |     |     |
| 28%              |     |     |     |     |     |     |
| 20%              |     |     |     |     |     |     |
| 24%              |     |     |     |     |     |     |
| 9%               |     |     |     |     |     |     |
| 18%              |     |     |     |     |     |     |
| -28%             |     |     |     |     |     |     |
| -5%              |     |     |     |     |     |     |
| -6%              |     |     |     |     |     |     |
| 11%              |     |     |     |     |     |     |
| 137%             |     |     |     |     |     |     |
| 58%              |     |     |     |     |     |     |
| 40%              |     |     |     |     |     |     |
| -15%             |     |     |     |     |     |     |
| 23%              |     |     |     |     |     |     |
| -15%             |     |     |     |     |     |     |

Figure 2:
Figure 3:
Figure legends

**Figure 1** Selection efficacy ($d_N/d_S$), polymorphism ($\hat{\theta}_w$ and $p_N/p_S$) and haploid genome size measurements for 11 pairs of surface and subterranean asellid species. Vertical bars next to the tree indicate species pairs with their surface (black circles) and subterranean (white circles) species. Numbers along branches of the tree are the numbers of single copy genes used to estimate the $d_N/d_S$. Color boxes indicate statistical support (p-value<0.05) in favor of (dark brown) or against (light brown) a decrease in selection efficacy or population size, or an increase in genome size in subterranean species. No box indicates no statistical differences between species of a pair. Error bars represent 95% bootstrap confidence intervals, except for genome size where it represents the range around the mean for 5 individuals. The percentage change from surface water to subterranean species is shown for each species pair.

**Figure 2** Variation in haploid genome size associated with the ecological transition from surface water to groundwater in 47 species, including isopods (top panel), decapods and gastropods (bottom panel). The 18 independent pairs of surface and subterranean species are delimited with boxes. Legends as in Figure 1. Identical species names followed by locality names within brackets refer to cryptic species (Morvan et al., 2013).

**Figure 3** Repeatome size estimates and composition using low coverage genome sequencing. A: size of the non-repeated genome (blue) and repeatome (orange) for the 22 species (tree symbols as in figure 1). B: Repeat family frequency spectrum. C: Number of shared repeats between two species as a function of divergence time. Relationship between GS and repeatome size (D), the number of repeat families (E) and the total genomic size of the 10 biggest repeat families (F). In the panel D, E and F, surface (black circles) and subterranean (white circles) species of a pair are joined by a grey line.
Table 1: Phylogenetic generalized least squares (PGLS) models testing the correlation between two variables. On the top of the table, correlation tests between the transition to groundwater (ecological status, proportion of subterranean branch, or colonization time) and variables ranging from selection efficacy ($dN/dS$, $w^-$), rate of adaptive evolution ($w^+$, $fq(w^+)$, $DoS$), polymorphism ($\hat{\theta}_w$ and $pN/pS$), and phenotypic traits (growth rate and body size) are reported. Another set of correlation tests between GS, or repeatome size, and some of these variables is also reported at the bottom of the table. Coefficients are in contrast to the surface status. Only comparable AIC are shown (same dependent variable and same number of observations). $w^-$: intensity of purifying selection, $w^+$: intensity of positive selection, $fq(w^+)$: frequency of sites under positive selection, $DoS$: direction of selection, LRT p-value: likelihood ratio test between the models with and without the given predictor variable. $R^2$: Cox and Snell generalized $R^2$. n: number of observations. *: p-value < 0.05, **: p-value < 0.01. The relative colonization time represents the proportion of the terminal branches estimated to be subterranean using the opsin gene, it is equal to $\text{time}_{\text{colonization}} / \text{time}_{\text{speciation}}$.

| Dependent variable | Predictor variable | n   | LRT p-value | Coefficient | AIC  | $R^2$ |
|--------------------|--------------------|-----|-------------|-------------|------|-------|
| $dN/dS$            |                    | 22  | 0.015*      | 0.017       | 0.237|
| $w^-$              |                    | 22  | 0.020*      | 0.018       | 0.217|
| $w^+$              |                    | 22  | 0.355       | -0.197      | 0.038|
| $fq(w^+)$          | ecological status  | 22  | 0.653       | 0.001       | 0.009|
| $DoS$              | ecological status  | 22  | 0.410       | 0.056       | 0.030|
| $\hat{\theta}_w$  | relative colonization time | 19 | 0.001** | 0.064 | 0.427 |
| $pN/pS$            | colonization time  | 19  | 0.016*      | 0.035       | 0.033|
| growth rate        | colonization time  | 19  | 0.486       | -0.001      | 0.025|
| body size          | colonization time  | 19  | 0.355       | 0.058       | 0.044|
|                    | ecological status  | 22  | 0.014*      | 0.340       | 29.1 |
|                    | colonization time  | 19  | 0.019*      | 0.789       | 25.0 |
| $dN/dS$            | relative colonization time | 22 | 0.009** | 10.447 | 28.3 |
| $w^-$              | relative colonization time | 22 | 0.004** | 9.897 | 26.8 |
| $w^+$              | relative colonization time | 19 | 0.521 | -0.465 | 0.021|
| $fq(w^+)$          | ecological status  | 19  | 0.016*      | 0.006       | 0.263|
| $DoS$              | ecological status  | 19  | 0.817       | 0.035       | 0.003|
| $\hat{\theta}_w$  | colonization time  | 19  | 0.486       | -0.001      | 0.025|
| $pN/pS$            | colonization time  | 19  | 0.355       | 0.058       | 0.044|
| growth rate        | colonization time  | 13  | 0.115       | -7.813      | 0.174|
| body size          | colonization time  | 19  | 0.581       | -1.134      | 0.016|
| Genome size        | ecological status  | 22  | 0.017*      | 0.251       | 29.1 |
|                    | colonization time  | 19  | 0.003**     | 0.724       | 37.2 |
| $dN/dS$            | ecological status  | 22  | 0.006**     | 8.189       | 15.3 |
| $w^-$              | ecological status  | 22  | 0.002**     | 7.808       | 13.5 |
| $\hat{\theta}_w$  | ecological status  | 22  | 0.437       | -50.457     | 22.2 |
| $pN/pS$            | ecological status  | 22  | 0.419       | 0.633       | 22.1 |
| growth rate        | ecological status  | 16  | 0.148       | -0.017      | 0.123|
| body size          | ecological status  | 22  | 0.916       | 0.003       | 22.8 |

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Table 2: Phylogenetic generalized least squares models testing the association between genome size (dependent variable) and the size and composition of the 22 species repeatomes. Same abbreviation as in Table 1. TGS: total genomic size, Gb: giba base, top1: most invasive repeat, orphan: repeat element found in one species only.

| Predictor variable                      | LRT p-value | Coefficient | AIC | R²  |
|-----------------------------------------|-------------|-------------|-----|-----|
| repeatome size (Gb)                     | <.0001***   | 1.295       | -33.0 | 0.955 |
| non-repeated genome size (Gb)           | <.0001***   | 2.773       | 6.5  | 0.728 |
| number of repeats                       | <.0001***   | 0.007       | 3.7  | 0.760 |
| top1 TGS (Gb)                           | 0.0022***   | 20.546      | 25.7 | 0.348 |
| top10 TGS (Gb)                          | <.0001***   | 5.407       | 1.6  | 0.783 |
| repeat orphans TGS (Gb)                 | 0.0163*     | 0.002       | 29.4 | 0.231 |
| number of orphan repeats                | 0.0056**    | 0.003       | 27.5 | 0.294 |
| orphans TGS / non-orphans TGS           | 0.3797      | -0.203      | 34.4 | 0.034 |