Fine Particulate Matter and Poor Cognitive Function among Chinese Older Adults: Evidence from a Community-Based, 12-Year Prospective Cohort Study

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BACKGROUND: Research on the relationship between long-term exposure to particulate matter with aerodynamic diameter ≤2.5 μm (PM2.5) and poor cognitive function is lacking in developing countries, especially in highly polluted areas.

OBJECTIVES: We evaluated associations of long-term exposure to PM2.5 with poor cognitive function in a diverse, national sample of older adults in China.

METHODS: This analysis included data on 13,324 older adults (5,879 who were 65–79 years of age, 3,052 who were 80–89 years of age, 2,634 who were 90–99 years of age, and 1,759 who were ≥100 years of age) with normal cognitive function at baseline from March 2002 to September 2014, with 64,648 person-years of follow-up. We used a geographic information system analysis to estimate the annual average satellite-derived PM2.5 concentration for the geocoded location of the participants’ baseline residences. Poor cognitive function was defined as a score of less than 18 on the Chinese version of the Mini-Mental State Examination (MMSE). Competing risk models were performed to explore the association of PM2.5 with poor cognitive function.

RESULTS: Each 10-μg/m³ increase in PM2.5 was associated with a 5.1% increased risk of poor cognitive function [adjusted hazard ratio (HR): 1.051; 95% confidence interval (CI): 1.023, 1.079]. Compared to the lowest quartile of PM2.5 (<41.4 μg/m³), adjusted HR values were 1.20 (95% CI: 1.09, 1.33), 1.27 (95% CI: 1.15, 1.41), and 1.21 (95% CI: 1.09, 1.34) for the second (≥41.4–50.3 μg/m³), third (≥50.3–60.7 μg/m³), and fourth (≥60.7 μg/m³) quartiles of PM2.5, respectively (p for trend <0.001). Subgroup analyses suggested stronger associations between PM2.5 and poor cognitive impairment in men than women. The association was positive in the 65- to 79- and ≥100-y age group but not significant and positive in the other two age groups with similar results.

CONCLUSION: PM2.5 was identified as a risk factor for poor cognitive function in Chinese older adults. Improving air quality may reduce the future population burden of poor cognitive function, especially in areas with high air pollution. https://doi.org/10.1289/EHP5304

Introduction

The aging population is projected to continuously increase in developed and developing parts of the world (Gerland et al. 2014). With the decline in fertility and mortality rates, China has become a rapidly aging society. China’s population of 60 years of age and above was 249 million in 2018, accounting for 17.90% of the country’s total population; this group, in turn, accounts for about one-fifth of the world’s total aging population and half of Asia’s aging population (National Bureau of Statistics 2019). The physical health of older adults progressively declines with age. Cognitive decline is the single most feared aspect of aging old (Martin 2004). A systematic review of studies conducted in China between 2001–2016 estimated that the prevalence of mild cognitive impairment (MCI) was 14.7% (Xue et al. 2018). The burden of poor cognitive function is expected to increase with the rapid increase in the elderly population. Therefore, an improved understanding of risk factors for poor cognitive function is needed to inform future disease control and prevention efforts.

Older adults may be especially vulnerable to hazards in their immediate environment, including environmental pollutants (Gouveia and Fletcher 2000; Katsouyanni et al. 2001) and fine particulate matter air pollution specifically (Naghavi et al. 2015). There is emerging evidence for the association between exposure to air pollution and poorer brain health (clinical dementia, neuroimaging correlates, or cognitive impairment) (Russ et al. 2019). Most previous studies of air pollution and cognitive function in the elderly have conducted in developed countries with relatively low air pollution exposures, including cross-sectional studies (Ailshire and Crimmins 2014; Ailshire and Clarke 2015; Gatto et al. 2014; Ranft et al. 2009; Shin et al. 2019; Tzivian et al. 2016) and prospective cohort studies (Power et al. 2011; Tonne et al. 2014; Wellenius et al. 2012; Wu et al. 2012; Loop et al. 2013). Previous studies of air pollution and poor cognitive function among elderly Chinese participants have been limited to a cross-sectional study (Sun and Gu 2008; Zeng et al. 2014) and a repeated measures analysis (Zhang et al. 2018) that used the air pollution index (API), a simplified measure of air quality that reflects the concentrations of several air pollutants (sulfur dioxide, nitrogen dioxide, and inhalable particulates) to estimate exposure at the city or county level.

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In this study, we analyzed data from a large, prospective, wide-ranging representative cohort of Chinese adults ranging from 65–114 years of age at baseline. To test the hypothesis that long-term exposure to particulate matter with aerodynamic diameter \( \leq 2.5\ \mu m \) (PM\textsubscript{2.5}) is associated with poor cognitive function, competing risk models were performed, accounting for possible bias from selective attrition.

**Methods**

**Study Population**

This study is based on the Chinese Longitudinal Healthy Longevity Study (CLHLS), conducted in 866 highly diverse counties/cities in 23 provinces of China. Younger elderly (65–79 years of age) were first included in the CLHLS in 2002; in the present study, we included participants enrolled from 2002 to 2014. The CLHLS randomly selected half of the counties and cities in 23 of China’s 31 provinces; it was the first national longitudinal survey on determinants of healthy aging and includes the largest sample of adults \( \geq 80 \) years of age in China. Details of the study have been provided elsewhere (Lv et al. 2018). The study was approved by the Biomedical Ethics Committee of Peking University (IRB00001052-13074). All participants or their legal representatives signed written consent forms to participate in the baseline and follow-up surveys.

A total of 26,675 older adults were evaluated (standardized questionnaires and physical measurements performed) in the CLHLS from 2002 to 2014. A total of 13,324 adults \( \geq 65 \) years of age with normal cognition at enrollment were included in the present analysis, including 5,879 who were \( \geq 65 \) years of age with normal cognition at enrollment were included in the CLHLS in 2002; in the present study, we included participants enrolled from 2002 to 2014. The CLHLS randomly selected half of the counties and cities in 23 of China’s 31 provinces; it was the first national longitudinal survey on determinants of healthy aging and includes the largest sample of adults \( \geq 80 \) years of age in China. Details of the study have been provided elsewhere (Lv et al. 2018). The study was approved by the Biomedical Ethics Committee of Peking University (IRB00001052-13074). All participants or their legal representatives signed written consent forms to participate in the baseline and follow-up surveys.

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**Measurement of Exposure to PM\textsubscript{2.5}**

Nationwide PM\textsubscript{2.5} monitoring data were available beginning in 2013. Our exposure estimates were derived from a remote-sensing PM\textsubscript{2.5} concentration grid data product with a resolution of 0.01 × 0.01 degrees provided by the Institute of Atmospheric Physics, University of Dalhousie (van Donkelaar et al. 2016; Boys et al. 2014). This data set is the longest and highest-resolution–exposure data set available in China and widely used for evaluation of air pollution and health outcomes (Yin et al. 2017; Crouse et al. 2015). The data product used aerosol optical depth (AOD) to simulate PM\textsubscript{2.5} concentration with a land-use regression (LUR) model. Subsequently, the data set was calibrated based on global ground-based PM\textsubscript{2.5} observations using the geographically weighted regression (GWR) method. In addition, van Donkelaar’s team collected 210 global data sets of ground observations from the literature and used them to estimate PM\textsubscript{2.5} for global satellite inversion and found important consistency (\( R^2 = 0.81 \)), with research areas including northern India and eastern China (2001–2010) (van Donkelaar et al. 2015). Missing PM\textsubscript{2.5} concentration data may exist in some grids due to the unavailability of remote-sensing data because of the influence of clouds and snow-capped mountains. We therefore excluded the participants without valid PM\textsubscript{2.5} data at the stage of selecting the study population (\( n = 81 \)) (Figure 1).

PM\textsubscript{2.5} concentrations were assigned to the participants based on the following rules: \( a \) the household address of each surveyed person was obtained via questionnaire, including information on province, city, district (or county), and street (or village). Using R software (version 3.3.1; R Development Core Team), we matched each ascertained address to a series of latitude and longitude coordinates; \( b \) we used geographic information systems vector data to generate a basic map of China’s districts and counties to calculate and display PM\textsubscript{2.5} concentration. These vector data are publicly available from the National Geographic Information Center (http://www.webmap.cn/main.do?method=index); and \( c \) each residual address was subsequently matched with the grid of the ambient PM\textsubscript{2.5} model it fell into to assign PM\textsubscript{2.5} concentrations at a spatial resolution of 0.01 degrees.

In order to characterize the historic long-term exposures more accurately, we averaged the predicted PM\textsubscript{2.5} concentrations for all residential locations where individual participants lived from recruitment to a diagnosis of poor cognitive function, death, loss to follow-up, or the end of follow-up (September 2014). We also calculated exposures for the 3 y prior to a diagnosis of poor cognitive function, death, loss to follow-up, or the end of follow-up (September 2014) to explore whether different metrics of PM\textsubscript{2.5} exposure may alter the findings.

**Assessment of Cognitive Function**

Cognitive function was assessed by the Chinese version of the MMSE administered by a trained staff of the Center for Disease Control and Prevention of China or by a trained student. The MMSE has been widely applied in epidemiological studies as a screening test for poor cognitive function and to track changes in cognitive function over time (Daniels et al. 2011; van Exel et al. 2003). The MMSE is also used as a screen for dementia, with high specificity (usually above 0.80) and moderate reliability (24-h test–retest by Pearson correlation usually above 0.85) from a meta-analysis conducted in 2013 (Mitchell 2013). We used a version of the MMSE that has been adapted for the cultural and socioeconomic conditions in China (Zeng and Vaupel 2002). For the sample in the CLHLS 2002 wave, the reliability of the MMSE scale is high (Cronbach’s \( a = 0.96 \)) (Sun and Gu 2008). The assessment tool addresses the following aspects of cognitive functioning: orientation, registration, attention, memory, language, and visual construction skills. MMSE scores range from 0 to 30; a higher total score indicates better cognitive function. Because more than half of the participants were illiterate, we used a relatively low cutoff score of 18 to define poor cognitive function and used scores \( \geq 18 \) to define normal cognitive function (Tombaugh and McIntyre 1992; Zhang et al. 1990; Cui et al. 2011).

**Covariates**

Covariates were obtained using a structured face-to-face questionnaire administered by trained interviewers. Potential confounders included age (continuous), sex, urban or rural residence (city vs. town or countryside), current marital status (either married or divorced, widowed, or never married), living pattern (either with family members, living in a nursing home, or alone), education (literate or illiterate based on >1 or <1 y of formal education), smoking status (current smoker, former smoker, or never smoker, defined by the questions “Do you currently smoke?” and “Did you previously smoke?”), alcohol drinking status (current drinker, former drinker, or never drinker, based on the questions “Do you currently drink alcohol?” and “Did you previously drink alcohol?”), and regular exercise (“Do you exercise regularly?”). In addition, we collected information on medical history, including self-reported diagnoses of diabetes (yes or no), heart disease (yes or no), and respiratory disease (yes or no), hypertension (yes or no) based on self-report of a diagnosis by a doctor or physician, or on
measured systolic blood pressure $\geq 140$ mmHg or diastolic blood pressure $\geq 90$ mmHg), and disability in activities of daily living (ADL) (yes or no) (Han et al. 2019; Li et al. 2018; Tervo et al. 2004). Blood pressure was measured using a mercury sphygmomanometer after 5 min of rest, and the average value of the two measurements was used for analysis. Disability in ADL was defined as the inability to independently perform any of the following tasks: bathing, toileting, dressing, eating, continence, cleaning themselves afterward, or indoor movement (Katz et al. 1963). In addition, we used the gross domestic product (GDP) and number of physicians per 1,000 residents at the prefecture-level as indicators of socioeconomic status.

**Statistical Analysis**

The differences of categorical variables were tested with the Cochran-Armitage test for trends; the differences of continuous variables were tested with analysis of variance among participants categorized by quartiles of PM$_{2.5}$ exposure. Multiple imputation (MI) is an attractive approach for missing data problems. To impute missing values of covariates, proc MI in SAS (version 9.4; SAS Institute, Inc.) was performed with the Markov chain Monte Carlo missing pattern. For a continuous variable, a regression method was used; for a classification variable, logistic regression method was used when the classification variable had a categorical response (dichotomous and polytomous variables). An MI procedure (five imputations) replaces each missing value of covariates ($<1\%$ missing for each individual characteristic), so that all 13,324 participants with outcomes of cognitive function were included (Yuan 2011).

Associations between average annual PM$_{2.5}$ exposures and the incidence of poor cognitive function were estimated using Cox proportional hazard models [PROC PHREG in SAS (version 9.4; SAS Institute, Inc.)], with follow-up time ending at the date of diagnosis or censored at the date of death (since death is a competing risk), loss to follow-up, or the end of follow-up. The date of death was confirmed by the participant’s closest relatives or the village doctor. Covariates included age, sex, urban/rural residence, marital status, education, living pattern, smoking, alcohol drinking, regular exercise, diabetes, heart disease, hypertension, respiratory disease, disability in ADL, and prefecture-level GDP and physicians per 1,000 residents. PM$_{2.5}$ was categorized by quartiles ($<41.4$, $\geq 41.4–50.3$, $\geq 50.3–60.7$, and $\geq 60.7$ $\mu g/m^3$) or according to Chinese guidelines ($<15$, $\geq 15–35$, $\geq 35–75$, $\geq 75$ $\mu g/m^3$) (MEP...
Subgroup analyses were conducted to assess whether associations with a 10-μg/m³ increase in PM$_{2.5}$ varied by age group (65–79 y, 80–89 y, 90–99 y, ≥100 y), sex, education (literate or illiterate), residence (urban or rural), smoking status (current vs. former or never), drinking status (current vs. former or never), regular exercise (yes or no), comorbidity (yes or no), disability in ADL (yes or no), region (east, central, or west China), and prefecture-level GDP by tertiles (<10,844.52, ≥10,844.52–22,159.52, and ≥22,159.52). To explore differences in hazard ratios between/among subgroups, we also calculated p-values for interaction. Statistical interaction was performed for building Cox models, including two interaction terms and their product. The maximum likelihood ratio test was used to detect the significance of the interaction effect. For the polytomous variables such as age, region, and GDP, ordinal variables were included to estimate the interaction in the Cox model. In the subgroup analyses, separate stratum-specific models were performed to derive the hazard ratios for each category without interaction terms.

To evaluate the robustness of our estimates, several additional sensitivity analyses were performed. Since loss to follow-up may have been more likely to occur in participants at higher risk of poor cognitive function or death, analyses were conducted excluding the participants lost to follow-up. To assess whether overall results were affected by the inclusion of people who were at high risk of adverse outcomes concurrently with short-term exposure of PM$_{2.5}$, models were repeated with exclusion of participants who died in the first year. We also report results from an unadjusted model and a model adjusted for age only.

PM$_{2.5}$ exposure matching was completed using ArcGIS (version 9.3; Esri) and R (version 3.3.1; R Development Core Team) software. All statistical analyses were conducted using SAS (version 9.4; SAS Institute, Inc.) and R (version 3.3.1; R Development Core Team). Two-sided p-values (p < 0.05) were considered statistically significant.

Results
The average age of the study sample at baseline was 82.4 ± 11.9 y, ranging from 65 to 114 y; 47.5% were men, 18.8% lived in urban areas, 41.4% were married, and 58.5% were illiterate (Table 1). Men had a higher percentage of smoking and alcohol drinking than women (Table S1).

Average annual PM$_{2.5}$ exposures over the follow-up period ranged from 8.5 to 110.7 μg/m³ [median: 50.1 μg/m³; interquartile range (IQR): 19.5 μg/m³]. There was substantial variability in the average annual exposures across the study area (Figure 2). Average annual PM$_{2.5}$ exposures during each year of the study period (2002–2014) were highly correlated (pairwise correlation coefficients: 0.84–0.98) (Table S2). The distributions of PM$_{2.5}$ exposures among participants followed an approximately normal distribution each year (Figure S1).

A total of 3,271 participants who had normal cognitive function at baseline (MMSE score ≥18) developed poor cognitive function during 64,648 person-years of follow-up. During the follow-up, 7,458 deaths were documented (including 1,464 with poor cognitive function). Based on adjusted competing risk models, a 10-μg/m³ increase in PM$_{2.5}$ was associated with a 5.1% increase in the risk of poor cognitive function [hazard ratio (HR): 1.051 (95% CI: 1.023, 1.079)] (Figure 3). Corresponding estimates from a crude model and a model adjusted for age only were an HR of 1.044 (95% CI: 1.017, 1.071) and an HR of 1.054 (95% CI: 1.028, 1.081), respectively (Table S3). In contrast, the HR for all-cause mortality with each 10-μg/m³ increase in PM$_{2.5}$ was 1.007 (95% CI: 0.998, 1.026).

Compared to the lowest quintile of PM$_{2.5}$ (<41.4 μg/m³), the adjusted HRs for ≥41.4–50.3 μg/m³, ≥50.3–60.7 μg/m³, and ≥60.7 μg/m³ were 1.20 (95% CI: 1.09, 1.33), 1.27 (95% CI: 1.15, 1.41), and 1.21 (95% CI: 1.09, 1.34), respectively (p for trend <0.001). When categorized by Chinese guidelines, adjusted HRs were 0.50 (95% CI: 0.12, 2.00), 1.20 (95% CI: 1.09, 1.34), and 1.28 (95% CI: 1.04, 1.59) for <15, ≥15–35, and ≥35 μg/m³ PM$_{2.5}$ levels, respectively, compared with ≥15–35 μg/m³ (p for trend <0.001) (Figure 3). Compared to the lowest quintile of PM$_{2.5}$ (<41.4 μg/m³), the crude models’ HRs were ≥41.4–50.3 μg/m³, ≥50.3–60.7 μg/m³, and ≥60.7 μg/m³ were 1.12 (95% CI: 1.02, 1.24), 1.23 (95% CI: 1.12, 1.36), and 1.29 (95% CI: 1.17, 1.42), and the models adjusted for age-only HRs were 1.19 (95% CI: 1.08, 1.31), 1.29 (95% CI: 1.17, 1.42), and 1.24 (95% CI: 1.13, 1.37), respectively (Table S3).

The association between poor cognitive function and PM$_{2.5}$ during the 3 y prior to the end of each individual’s follow-up was positive but weaker than associations with average exposure over the entire follow-up period: the adjusted HR for a 10-μg/m³ increase in PM$_{2.5}$ was 1.033 (95% CI: 1.007, 1.058) (Figure 3). When PM$_{2.5}$ was modeled using penalized splines, the association with the ln(HR) for poor cognitive function was approximately linear; with the increment of PM$_{2.5}$, the risk of poor cognitive function significantly increased (Figure 4).

Adusted associations were consistent with the primary model estimates after excluding 1,423 participants lost to follow-up (HR: 1.062; 95% CI: 1.034, 1.089 for a 10-μg/m³ increase in PM$_{2.5}$) and after excluding 865 participants who died during the first year of follow-up (HR: 1.056; 95% CI: 1.029, 1.084) (Table S4).

In subgroup analyses, the association between a 10-μg/m³ increase in PM$_{2.5}$ and poor cognitive function differed by sex, with a stronger association among men (HR: 1.073; 95% CI: 1.028, 1.120) than women (HR: 1.036; 95% CI: 1.001, 1.073) (p-interaction = 0.06). When evaluated by region, associations were very similar for eastern and central China (HR: 1.059; 95% CI: 1.021, 1.097 and HR: 1.058; 95% CI: 1.011, 1.108, respectively), while the association was stronger for western China (HR: 1.151; 95% CI: 0.967, 1.369) but much less precise, consistent with the smaller numbers of events and participants in this region (p-interaction = 0.15). The association between poor cognitive function and a 10-μg/m³ increase in PM$_{2.5}$ increased as prefecture-level GDP decreased, with HRs of 1.025 (95% CI: 0.983, 1.068), 1.052 (95% CI: 1.003, 1.103), and 1.081 (95% CI: 1.026, 1.138) for the upper, middle, and lower tertiles, respectively (p-interaction = 0.16), with a slight inverse association for the relatively small number of participants classified as disabled (HR: 0.991; 95% CI: 0.918, 1.069) and a positive association for nondisabled participants (HR: 1.061; 95% CI: 1.031, 1.092, p-interaction = 0.29). The HR for those with comorbidities is stronger than for those without (HR: 1.078; 95% CI: 0.010, 1.151 and HR: 1.046; 95% CI: 1.105, 1.077, respectively), but the difference is not significant (p-interaction = 0.73). For smoking, the association was stronger for current smokers than for former or never smokers, though the difference was not significant (p-interaction = 0.24). For age, the association was weakest for the ≥100-y group and similar for the three younger age groups. The association was essentially null for urban residents (HR: 1.008; 95% CI: 0.951, 1.069) but positive for rural residents (HR: 1.061; 95% CI: 1.029, 1.094), though
the differences were not significant, in part due to the small numbers of participants in urban areas (p-interaction = 0.23) (Figure 5).

Discussion

In this prospective cohort study based on a nationwide sample of older adults living in 866 highly diverse counties and cities of China, including areas exposed to high concentrations of PM$_{2.5}$, exposure to PM$_{2.5}$ (8.5–110.7 μg/m$^3$) was associated with a higher risk of poor cognitive function after accounting for potential confounders.

Studies of the impact of PM$_{2.5}$ on cognitive function in elderly populations living in developed countries (the United States and United Kingdom) have also reported evidence of an adverse effect (Wuve et al. 2012; Wellein et al. 2012; Tonne et al. 2014; Ailshire and Clarke 2015). Loop et al. (2013) did not find clear evidence of an association between PM$_{2.5}$ and the incidence of cognitive impairment in a geographically diverse, biracial U.S. cohort of men and women (n = 20,150), with an odds ratio (OR) = 1.26 (95% CI: 0.97, 1.64) for increased odds of incident impairment with a 10-μg/m$^3$ increase in PM$_{2.5}$ concentration. The OR was attenuated toward 1 after adding more covariates. The highest quartile of PM$_{2.5}$ concentration was 14.8–21.0 μg/m$^3$ in Loop et al. (2013)’s study, which would have fallen within the lowest quartile for our study (<4.1 μg/m$^3$). Compared to other prospective
cohort studies focusing on PM$_{2.5}$ (Tonne et al. 2014; Weuve et al. 2012), the exposure data reported by Loop et al. (2013) are similar to these two studies. Tzivian et al. (2016) reported positive cross-sectional associations between long-term air pollutant exposures and MCI during 5 y of follow-up in a cohort of German adults (45–75 years of age at baseline), with an OR of 1.16 (95% CI: 1.05, 1.27) for overall MCI in association with an IQR increase in PM$_{2.5}$ (1.44 µg/m$^3$) and similar or slightly weaker associations with IQR increases in long-term PM with aerodynamic diameter ≤10 µm, PM$_{2.5}$ absorbance, NO$_2$, and NO$_x$ concentrations. The results of our study provide further support the potential influence of long-term exposure to PM$_{2.5}$ on the development of poor cognitive function in a population with PM$_{2.5}$ exposure concentrations ranging from relatively low values to concentrations that are much higher than those observed in developed countries. In China, previous studies have explored the role of air pollution in global cognitive performance (Sun and Gu 2008; Zeng et al. 2014; Zhang et al. 2018). Sun and Gu (2008) reported that a 1-point increase in API is associated with cognitive function test score (linear coefficient = −1.51; 95% CI: −2.16, −0.86). Zeng et al. (2014) explored the relationship between API and cognitive dysfunction in the CLHLS study; each unit increase in API was associated with a 9% increased odds of age-related cognitive decline. Zhang et al. (2018) found that long-term exposure to air pollution impedes cognitive performance, with a 1-standard deviation increase in average 3-y API associated with a 1.132-point drop in cognitive function test score.

Table 1. The association of particulate matter with aerodynamic diameter ≤2.5 µm (PM$_{2.5}$) and poor cognitive function [Mini-Mental State Examination (MMSE) score <18] in competing risk models among Chinese older adults 65 years of age and older. Adjusted covariates include age (continuous), sex, residence, current marital status, living pattern, education (literacy status), smoking status, alcohol drinking status, regular exercise, diabetes, heart disease, hypertension, respiratory disease, disability in activities of daily living, gross domestic product (GDP), physicians per z persons at prefecture level. Note: CI, confidence interval; HR, hazard ratio.

Figure 2. Map of particulate matter with aerodynamic diameter ≤2.5 µm (PM$_{2.5}$) concentration in China during the study period (2002–2014). The blue lines denote the geographic areas included in this study, covering 23 provinces from the south to the north of China. Black points represent the location of participants. The inset in the lower right corner indicates the South China Sea and the islands. Our large study area included a large sample size and encompassed a broad concentration range of fine particulate matter that included areas of the lowest and highest pollution in China.
in verbal test scores. The API includes values for various contaminants, making it difficult to attribute their findings to a specific pollutant or set of pollutants. Our findings complement and extend these previous findings by estimating the risk of poor cognitive function in association with ambient PM$_{2.5}$ exposures in a large population of elderly Chinese. PM$_{2.5}$ contains a variety of components. The main components are elemental carbon, organic carbon compounds, sulfates, nitrates, ammonium salts, and various metal elements. Although we found an association between PM$_{2.5}$ and poor cognitive function, we could not determine whether specific components of PM$_{2.5}$ were primarily responsible. In addition, we cannot rule out confounding by other air pollutants.

As observed in previous studies (Ailshire and Clarke 2015; Shin et al. 2019), subgroup analyses suggested stronger associations between PM$_{2.5}$ and cognitive impairment in men than women. Men in our study were more likely than women to be current smokers (39% vs. 7.4%), alcohol consumers (36% vs. 11%), and regular exercisers (38% vs. 27%). Differences in gender-related lifestyle factors may result in differential exposure patterns between men and women and may modify the effects of environmental exposures. The association between PM$_{2.5}$ and poor cognitive function was stronger among residents of rural areas, though the difference was not significant. This could reflect a greater susceptibility among rural residents, differences in ambient air pollutants between rural and urban areas, confounding (e.g., by exposure to indoor air pollution), or other mechanisms.

Epidemiological studies have shown that poor cognitive function in old age is a multifactorial disease (Avila et al. 2010; del Valle 2011). However, the specific relevance or causative nature of these factors and the potential for independent or synergistic effects of air pollution on the causes and pathogenesis of cognitive decline remain unclear. Atmospheric particulate matter inhaled through the respiratory tract activates human macrophages and inflammatory cytokines, causing inflammation and subsequent oxidative stress. Inflammatory compounds spread through the body’s circulatory system and penetrate the blood–brain barrier, thus potentially affecting nervous system function (Genc et al. 2012; Hirano et al. 2003). A second potential mechanism is related to the direct entry of PM into the nervous system through the olfactory bulb following nasal inhalation. Pathological changes in the olfactory bulb were observed in the early stages of Alzheimer’s disease (AD) (Doty 2008). Ultrafine particles (UFP) were observed in human olfactory bulb periglomerular neurons after exposure to air pollution (Calderón-Garciduenas et al. 2004).

Our research has the following strengths. First, based on the prospective cohort design, we were able to identify individuals with normal cognition at baseline to assess poor cognitive function in relationship to PM$_{2.5}$. Second, our study also provided an opportunity to explore the exposure–response relationship between PM$_{2.5}$ and poor cognitive function over a wide range of PM$_{2.5}$ concentrations (from 8.5 to 110.7 μg/m$^3$). Third, the large sample size of community-based older adults, especially ≥80 years of age, permitted us to focus not only on the association of PM$_{2.5}$ with poor cognitive function but also to estimate associations in different subgroups. The association was weakest and not significant for the ≥100-y group, positive in the 65–79 age group but not significant and positive in the other two age groups with similar results. Finally, our cohort contained a wide geographical distribution of residents that covered nearly all densely populated areas in China.

There are some limitations of the present study worth noting. First, the duration of residence at the current address for each study participant was unknown. We averaged exposures over all residences during the follow-up period, and because we were unable to weigh exposures according to the duration of residence in each location, some exposure misclassification would have occurred. According to the home address obtained in each follow-up visit, we have roughly counted that 356 of the 13,324 respondents changed their addresses and moved 1 times. Older adults (chosen for this study) tend to maintain geographic stability. Second, only outdoor exposures to PM$_{2.5}$ were estimated based on residential location; indoor air pollution and secondhand smoking, which may play a potential role in confounding the association of ambient PM$_{2.5}$ with poor cognitive function, were not measured. Third, only PM$_{2.5}$ was measured; other ambient pollutants were not assessed due to a lack of these data for China. There is a possibility that the effects observed in the present study are not due to PM$_{2.5}$ but to another pollutant or combination of pollutants whose concentrations strongly correlate with that of PM$_{2.5}$. Fourth, the simulation of the exposure data did not separately calibrate the model based on the values of the ground monitoring stations in different regions, but cross-validation was performed at a later stage. Fifth, the covariates or confounding factors we have collected may not be comprehensive or accurate; for example, in the definition of hypertension, we cannot completely exclude the possibility of individual blood pressure fluctuations, although we used reliable measurement methods. In addition, we did not have access to data about the number of cigarettes consumed, only whether the participant was a current, former, or never smoker. The smoking and drinking information in this study were self-reported, so they are subject to recall bias. Finally, the large portion of participants who died or were lost to follow-up may have led to selection bias. However, the competitive risk models and sensitivity analyses helped address this issue and provide evidence in support of an association of PM$_{2.5}$ and poor cognitive function.

Conclusions

In this community-based prospective cohort study among older adults in China, long-term exposure to PM$_{2.5}$ was associated with an increased risk of poor cognitive function. Overall, our findings...
suggest that reducing air pollution exposure may delay or prevent poor cognitive function in the population as a whole. This is of great significance to the economic and social development of China’s aging society. Future studies with longer periods of follow-up are needed to focus more comprehensively on the multiple components of air pollution and specific areas of cognition to better explore causality.

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Figure 5. Stratified analyses of association of each 10-μg/m³ increase in particulate matter with aerodynamic diameter ≤2.5 μm (PM2.5) concentration with poor cognitive function (Mini-Mental State Examination (MMSE) score <18) in competing risk models among Chinese older adults of 65 years of age and older. Adjusted covariates include age (continuous), sex, residence, current marital status, living pattern, education (literacy status), smoking status, alcohol drinking status, regular exercise, diabetes, heart disease, hypertension, respiratory disease, disability in activities of daily living, gross domestic product (GDP), physicians per 100 persons at the prefecture level. Note: CI, confidence interval; HR, hazard ratio.

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