**Research Paper**

**Early-life mitochondrial DNA damage results in lifelong deficits in energy production mediated by redox signaling in *Caenorhabditis elegans***

Kathleen A. Hershberger, John P. Rooney, Elena A. Turner, Lauren J. Donoghue, Rakesh Bodhicharla, Laura L. Maurer, Ian T. Ryde, Jina J. Kim, Rashmi Joglekar, Jonathan D. Hibshman, Latasha L. Smith, Dhaval P. Bhatt, Olga R. Ilkayeva, Matthew D. Hirschey, Joel N. Meyer

---

**Abstract**

The consequences of damage to the mitochondrial genome (mtDNA) are poorly understood, although mtDNA is more susceptible to damage resulting from some genotoxicants than nuclear DNA (nucDNA), and many environmental toxicants target the mitochondria. Reports from the toxicological literature suggest that exposure to early-life mitochondrial damage could lead to deleterious consequences later in life (the "Developmental Origins of Health and Disease" paradigm), but reports from other fields often report beneficial ("mitohormetic") responses to such damage. Here, we tested the effects of low (causing no change in lifespan) levels of ultraviolet C (UVC)-induced, irreparable mtDNA damage during early development in *Caenorhabditis elegans*. This exposure led to life-long reductions in mtDNA copy number and steady-state ATP levels, accompanied by increased oxygen consumption and altered metabolite profiles, suggesting inefficient mitochondrial function. Exposed nematodes were also developmentally delayed, reached smaller adult size, and were rendered more susceptible to subsequent exposure to chemical mitotoxicants. Metabolomic and genetic analysis of key signaling and metabolic pathways supported redox and mitochondrial stress-response signaling during early development as a mechanism for establishing these persistent alterations. Our results highlight the importance of early-life exposures to environmental pollutants, especially in the context of exposure to chemicals that target mitochondria.

---

**1. Introduction**

The mitochondrial genome (mtDNA) is small (16,569 bases in humans) and encodes a small number of genes: 13 proteins, 22 tRNAs and 2 rRNAs in humans, with similar numbers in most metazoans. Nonetheless, it is critical: the proteins are essential subunits of the mitochondrial respiratory chain (MRC) complexes, while the tRNAs and rRNAs are required for translation of those proteins. The health significance of mtDNA homeostasis is demonstrated by the fact that a large number of diseases are caused by mutations [1] or depletion [2] of mtDNA, which is normally present in thousands of copies per cell. In addition to these specific mtDNA-related diseases, mitochondrial dysfunction has been implicated in many more common human diseases including cancer, diabetes, metabolic syndrome and neurodegenerative conditions [3]. Further support for the importance of mitochondrial function to health is provided by evidence for toxicity associated with many pharmaceuticals that target mitochondria [4].

Globally, environmental pollutants are a major environmental driver of loss of years of life, conservatively estimated at three times greater than malaria, AIDS, and tuberculosis combined [5]. Developmental and low-level exposures are of particular concern [6,7]. The evidence that mitochondria are important targets of many environmental toxicants is growing, yet the consequences of such mitotoxicity are poorly understood [8]. mtDNA may be a particularly important target, because of the lack of some DNA repair pathways, in particular nucleotide excision repair (NER) [9]. Without NER, DNA lesions caused by common...
environmental stressors such as ultraviolet radiation, polycyclic aromatic hydrocarbons, and mycotoxins are irreparable in mtDNA. In vitro, such lesions can impair the progression of the mitochondrial DNA polymerase \( \gamma \) \(^{[10,11]} \), potentially resulting in decreased mtDNA copy number or mutagenesis in vivo. However, despite the fact that many common environmental exposures cause irreparable mtDNA damage \(^{[12]} \), the in vivo effects of such damage remain poorly understood.

We have previously employed ultraviolet C radiation (UVC) exposure to generate mtDNA damage that is irreparable due to the absence of NER (nucDNA damage is also caused, but is efficiently repaired). UVC is not environmentally relevant because stratospheric ozone absorbs UVC wavelengths; however, it provides a useful laboratory tool for near-exclusive production of damage (photodimers) that is repaired by NER, compared to UVB and UVA which also produce significant oxidative DNA damage. We reported that early-life (1st larval stage) exposures to such damage reduced levels of mtDNA, mtRNAs, total steady-state ATP levels, and oxygen consumption in developing C. elegans \(^{[13,14]} \). What might the long-term impact of such mtDNA damage be? Although irreparable, UVC-induced mtDNA damage is slowly removed via a process involving autophagic machinery in both adult C. elegans and human cells in culture \(^{[13,15]} \). We also found that the nematode mitochondrial DNA polymerase \( \gamma \), polg-1, was dramatically upregulated by UVC exposure \(^{[14]} \), suggesting the potential for an adaptive response. An adaptive response appears to be consistent with multiple examples of developmental mitochondrial dysfunction resulting in beneficial later-life outcomes in C. elegans and other species, including stress resistance and increased lifespan, sometimes termed “mitohormesis” \(^{[16–20]} \). On the other hand, a growing body of environmental health literature suggests that environmental exposures during development can lead to adverse outcomes later in life, a paradigm sometimes named the Developmental Origins of Health and Disease (DOHaD) \(^{[21,22]} \). A recent example identified long-term mitochondrial dysfunction associated with developmental mitochondrial dysfunction resulting from exposure to the Complex I inhibitor rotenone \(^{[23]} \).

Here, we asked if a low-level, in vivo, developmental environmental exposure that causes irreparable mtDNA damage would result in hormetic or adverse later life outcomes related to mitochondrial function, and tested the mechanisms mediating observed changes.

### 2. Results

We employed a previously developed \(^{[13,14]} \) protocol that involves exposing growth-arrested L1 larvae to 7.5 J/m\(^2\) UVC three times (Fig. 1A), 24 h apart, in order to cause a high level of irreparable mtDNA damage while permitting nucDNA repair. Importantly, this is not a highly toxic level of UVC exposure, as demonstrated by the fact that there was no effect on lifespan (Fig. 1B). However, this protocol consistently leads to a significant reduction in N2 (wild-type strain) worm length that is maintained at least 72 h post-UVC exposure (Fig. 1C). We then examined mitochondrial and organismal phenotypes at later timepoints, and tested signaling pathways involved in the response to this damage. Many of these studies were conducted using strains (gp-1 and PE255, an ATP reporter strain created in a glp-4 background) in which germline proliferation fails at 25 °C \(^{[24]} \), in order to exclude the confounding effects of the high level of mtDNA replication that occurs during gametogenesis \(^{[25]} \).

#### 2.1. mtDNA damage accumulates over consecutive UVC exposures

We first tested the persistence of UVC-induced mtDNA damage during the worms’ lifespan. UVC-induced DNA damage largely comprises photodimers and 6,4-photoproducts, both of which consist of...
covalent bonds between adjacent bases; we use the term “photolesion” in this manuscript to describe this class of DNA damage. Although photolesions are not repaired in mtDNA, they are slowly and incompletely removed by autophagy in adult nematodes [13]; furthermore, damaged genomes may be diluted by the extensive mtDNA replication that occurs during development [25]. We measured significantly higher levels of mtDNA damage immediately after the third and final dose of UVC exposure (time 0 days) compared to the non-exposed controls at the same time point (Fig. 1D and Fig. S1A). mtDNA damage in UVC-exposed animals at day 2 was significantly higher than in their non-exposed controls. All other data points showed no distinguishable differences between amount of mtDNA damage in control and UVC-exposed animals at the same time point. This suggests that mtDNA damage is either removed rapidly via mechanisms of mitochondrial quality control, which seems unlikely given the slow kinetics of photodimer removal [13], or that the mtDNA damage is diluted due to rapid proliferation of mitochondria during development [14]. The apparent (though not significant) slight increase in mtDNA damage in UVC-exposed animals at day 0 could be a result of oxidative damage resulting from mitochondrial reactive oxygen species (ROS) production; however, this assay cannot distinguish between different types of DNA damage. A significant increase in nucDNA lesions in UVC-exposed compared to control animals was observed at day 0, but not at any other time point. (Fig.1E and Fig. S1B). Notably, there was approximately 3-fold higher mtDNA damage in UVC-exposed animals at day 0 compared to nucDNA damage, indicating that nucDNA photolesion damage was removed rapidly between arrested L1 UVC exposures. This replicates our previous observations [13,14] of much faster photolesion removal in nucDNA, presumably largely from NER.

2.2. Reduction in mtDNA copy number is observed throughout life following developmental UVC exposure

To further characterize DNA integrity following UVC exposure, we analyzed mitochondrial and nuclear DNA copy number. We previously found that UVC exposure induced an 8-fold increase in polg-1 (mitochondrial DNA polymerase γ) expression 48 h after UVC-exposure [14]. We confirmed that polg-1 expression shows a strong trend of remaining elevated up to 4 days post-exposure (p = 0.11) and returns to control levels of expression by 8 days post-exposure (Fig. 1F). We hypothesized that photolesions in mtDNA would block the progress of POLG-1 [11], resulting in reduced mtDNA content at early timepoints post-exposure. However, because mtDNA damage is not significantly different in controls versus UVC-exposed worms in adulthood (Fig. 1D), we expected no difference in mtDNA copy number between groups in adulthood. Instead, mtDNA copy number per worm (Fig. S1C) and mtDNA/nucDNA ratio (Fig. 1G) was decreased throughout life in the UVC-exposed animals (p < 0.0001, main effect of UVC exposure). This was unexpected given the large proliferation of mtDNA copies during development and absence of detectable mtDNA damage in adults, and suggests that there could be other persistent effects of developmental exposure to UVC. There was no effect on nucDNA copy number (Fig. S1D). We observed similar results for the PE255 gfp-4(bn2) strain (Fig. S1E), used to permit comparisons to ATP levels (below).

2.3. Developmental mtDNA damage did not result in detectable changes in mtRNA transcription in adults

We next asked whether this reduction in mtDNA copy number, and/or the presence of undetected photolesions that might block the mtRNA polymerase [26], would reduce the levels of RNAs for mitochondrial proteins, or affect the stoichiometry of MRC proteins. All MRC complexes except complex II contain proteins coded in both the mitochondrial and nuclear genomes, and proper stoichiometry depends on coordinated production of proteins from both genomes. Improper stoichiometry results in protective mitochondrial unfolded protein responses (UPRmt) [27], which can be triggered by reduced transcription of mtDNA-encoded MRC subunits [17]. To our surprise, mtDNA-encoded transcript levels in UVC-treated nematodes were unchanged at 4- and 8-days post-exposure (Fig. S2A). This lack of decreased transcription in the UVC-exposed animals might result from the multiplicity of mtDNAs buffering against transcriptional blocks present in a subset of mtDNAs, possibly in combination with compensatory increased transcription. However, a series of experiments designed to detect a compensatory transcriptional response in nucDNA-encoded MRC subunits and the TFAM homologue hmg-5 indicated that there were again no differences in transcription between control and UVC-exposed animals (Figs. S2B-C). Additionally, we tested for, but did not detect, transcriptional activation of the UPRmt (Fig. S2D). The afs-1 mutant (which is deficient in mounting a UPRmt) was protected from larval growth inhibition (Fig. S2E), suggesting ATFS-1 mediated transcriptional response is important for inducing growth delay following UVC exposure. Finally, we asked whether exposure to chloramphenicol and doxycycline, two chemicals that inhibit mitochondrial protein translation and nematode larval development [28], would exacerbate UVC-mediated inhibition of larval growth [13]. We observed that both chloramphenicol and doxycycline further inhibited larval development after UVC exposure (Fig. S2F), indicating that, by itself, UVC-induced mtDNA damage does not entirely abrogate translation of proteins coded in mtDNA, even at the relatively high levels of mtDNA damage persisting in the larval stages [14]. Overall, this data suggests that mtDNA damage via UVC exposure does not result in overall shifts in transcriptional response of the mitochondrial or nuclear genomes, and that mitochondrial protein translation remains functional.

2.4. Mitochondrial respiration was altered in response to developmental mtDNA damage

Given that mtDNA copy number was reduced, but RNAs of mitochondrial proteins were expressed normally after UVC exposure, we next asked if mitochondrial function in later life was impacted by an early life exposure to UVC. To directly assess mitochondrial function, we used the Seahorse XF Analyzer to measure basal, ATP-linked, maximal, and non-mitochondrial respiration rates at days 4- and 8- post-UVC in vivo [29]. Unexpectedly, basal oxygen consumption rates were significantly increased in UVC-exposed nematodes at 4 days post exposure, but not at 8 days post-exposure (Fig. 2A). Basal respiration at 8 days post-exposure was significantly reduced in both the control and UVC exposed groups compared to basal respiration at 4 days post-exposure, consistent with decreased basal respiration with increased age. At 4 days post-exposure, the mitochondrial uncoupler FCCP increased oxygen consumption in non-exposed animals 2.2-fold, but had no effect in UVC-exposed nematodes (Fig. 2B). Surprisingly, ATP-dependent OCR measured by inhibiting ATP synthase with DCCD was significantly increased in UVC-exposed animals, compared to non-exposed animals at 4 days post-exposure (Fig. 2C). Spare capacity (the difference of maximal and basal respiration) was significantly lower in UVC-exposed animals compared to non-exposed animals at 4 days post-exposure (Fig. 2D), indicating that UVC-exposed animals were operating at or near maximal respiration. Interestingly, non-mitochondrial respiration was similar in the exposed and non-exposed groups at 4 days post-exposure, but there was a significant increase in non-mitochondrial oxygen consumption in the UVC-exposed animals at 8 days post-exposure (Fig. 2E). While non-mitochondrial oxygen consumption was increased in the UVC-exposed animals, there was no detectable difference in proton leak (Fig. 2F), suggesting that mitochondrial membrane potential is likely unchanged between control and UVC-exposed animals. To better understand possible functional consequences of increased basal oxygen consumption rates in the 4-day post-UVC exposure animals, we measured steady-state ATP levels using two methods. ATP
levels in extracts from *glp-1* nematodes at 4- and 8- days post-UVC exposure showed a strong trend of reduction compared to their relative controls (Fig. 2G). Consistent with previous results [30], ATP content decreased in an age-dependent fashion by roughly 50% between days 4 and 8 in both control and UVC-exposed animals. We also measured ATP using the *in vivo* ATP reporter strain PE255 and found similar results of decreased steady-state ATP in UVC-exposed animals compared to control animals at 8 days post-exposure. F. No differences in proton leak were observed. For data shown in A–F: Data are presented as mean with error bars representing the standard error. Open circles are individual data points (2–4 independent experiments, n = 7–8 per group per experiment). Letters show which groups are significantly different (p ≤ 0.05). Two-way ANOVA with Tukey’s correction for multiple comparisons. A UVC exposure of 2 J/m² was used in panels C and F because these experiments were performed at a later date. This exposure level produced the same growth inhibition as characterized in the paper. G. UVC exposure leads to a trend of reduced steady-state ATP in *glp-1* (q244). Two-way ANOVA with Tukey’s correction for multiple comparisons. n = 3 per group. H. The ratio of ATP per oxygen consumed was calculated by dividing the average steady-state ATP value by the average basal respiration value for each group.

Fig. 2. Early-life mtDNA damage results in deficient mitochondrial respiration and decreased ATP per oxygen consumption. A. Basal respiratory rates were significantly increased in UVC-exposed nematodes at 4 days post-exposure, but not at 8 days. B. Maximal respiration was unchanged between control and UVC-treated groups but decreased with age. C. ATP-linked respiration rates were significantly increased in UVC-exposed nematodes at 4 days post-exposure. D. Spare capacity was significantly decreased in UVC-exposed animals 4 days after exposure and returned to control levels by 8 days post-exposure. E. Non mitochondrial oxygen consumption was significantly higher in UVC-exposed animals compared to control animals at 8 days post-exposure. F. No differences in proton leak were observed. For data shown in A–F: Data are presented as mean with error bars representing the standard error. Open circles are individual data points (2–4 independent experiments, n = 7–8 per group per experiment). Letters show which groups are significantly different (p ≤ 0.05). Two-way ANOVA with Tukey’s correction for multiple comparisons. A UVC exposure of 2 J/m² was used in panels C and F because these experiments were performed at a later date. This exposure level produced the same growth inhibition as characterized in the paper. G. UVC exposure leads to a trend of reduced steady-state ATP in *glp-1* (q244). Two-way ANOVA with Tukey’s correction for multiple comparisons. n = 3 per group. H. The ratio of ATP per oxygen consumed was calculated by dividing the average steady-state ATP value by the average basal respiration value for each group.

The increased basal oxygen consumption and lack of spare respiratory capacity combined with reduced ATP levels in UVC-exposed nematodes suggested persistently inefficient mitochondrial function. The decrease in spare respiratory capacity might be partially explained by the decrease in mitochondrial content suggested by decreased mtDNA...
content, or by decreased availability of substrate. We also note that although at 8 days post exposure basal oxygen consumption rates were no longer elevated in UVC-treated worms, ATP levels still trended lower, suggesting that mitochondrial efficiency was also reduced, resulting in a reduction in ATP levels per unit oxygen consumed (Fig. 2H).

### 2.5. Metabolic flexibility is critical to tolerance of early-life UVC exposure

We hypothesized that in response to inefficient MRC function, UVC-exposed nematodes would shift metabolism towards alternate metabolic pathways. We tested whether strains carrying mutations in two major regulators of energy metabolism (*aak-2*, activated upon increase in the AMP:ATP ratio [34] and *nhr-49*, a nuclear hormone receptor that regulates energy metabolism in response to mitochondrial impairment [35]) would be sensitized to UVC-induced inhibition of larval development [13]. *aak-2* (Fig. 3A) and *nhr-49* (Fig. 3B) mutant strains were both approximately two-fold more sensitive to early-life UVC exposure, supporting the importance of the nematodes’ capacity to regulate energy metabolism in response to early-life mtDNA damage. Importantly for extrapolating the results from mutant versus wild-type N2 studies to the outcomes observed in *gfp-1* animals used in other experiments in this manuscript, the growth decrease observed in the N2 strain was also observed, although to a lesser extent, in the *gfp-1* strain (Fig. 3C).

Given this genetic evidence for protective metabolic restructuring, we next tested for alterations in transcript levels of key intermediary metabolic enzymes that regulate TCA cycle metabolism (*pck-1*, but detected no changes at either 4 or 8 days post-exposure (Fig. 5A). However, the lack of altered transcriptional regulation of intermediary metabolic pathways does not rule out significant metabolic alterations, as there are many non-transcriptional mechanisms of enzyme activity regulation. To directly test for metabolic shifts, we next measured levels of organic acids, amino acids and acyl carnitines.

### 2.6. Targeted metabolomics supports NADPH stress and altered metabolic function in adults

We observed age-related decreases in the levels of most organic acids between days 4 and 8 in both control and UVC-exposed animals (Fig. 4A). In contrast, fumarate and malate appeared to be elevated in the UVC-exposed animals compared to non-exposed controls and statistical analysis revealed that fumarate was significantly elevated in the UVC-exposed animals compared to controls, though statistical analysis revealed that fumarate was significantly elevated in the UVC-exposed animals compared to controls at 4 days post-exposure. Additional analyses including Western blots and fluorescence microscopy failed to see significant changes in key metabolic enzymes or enzyme activity.

### 2.7. Mitochondrial SOD mutants are more sensitive to UVC exposure

The data shown thus far provides evidence for NADPH stress as a mechanism of later life effects of early-life exposure to UVC. Specifically, we observed increased basal oxygen consumption (in the context of reduced ATP and movement) and evidence of increased flux through the malate-aspartate shuttle. Mitochondrial dysfunction, particularly in the context of increased overall oxygen consumption, can result in increased oxidative protein damage. We observed a significant increase in carbonyl groups of cytoplasmic (8%) and mitochondrial proteins (20%) 4 days post-UVC exposure (Fig. 5A and B), suggesting increased ROS production later in life in response to early-life mtDNA damage.

The ETC is a major source of intracellular ROS, and mitochondrial dysfunction can result in increased ROS production [38], including both superoxide anion (O$_2^-$) and hydrogen peroxide (H$_2$O$_2$). Mitochondrial ROS may cause damage or serve as mitochondrial signaling molecules [39]. To investigate the potential functional importance of increased mitochondrial ROS generation, we tested the sensitivity to developmental UVC exposure of nematode strains carrying loss-of-function mutations in all 5 superoxide dismutase (SOD) enzymes (*sod-1*, *sod-2*, *sod-3*, *sod-4*, and *sod-5*), which convert O$_2^-$ to H$_2$O$_2$ [40]. We additionally looked at the effects of UVC exposure on the double *sod-2,sod-3* mutant (mutations in both mitochondrial SOD proteins), and the triple *sod-1,sod-4,sod-5* mutant (mutations in all cytosolic and extracellular SOD proteins).

UVC-induced larval delay was exacerbated in both mitochondrial SOD mutants (*sod-2* and *sod-3*), with the delay occurring as soon as 48 h.
post-exposure in sod-2 nematodes and by 96 h post-exposure in sod-3 nematodes (Fig. 5C). Furthermore, the double sod-2;sod-3 mutant was significantly more growth delayed than wild-type 72 h after UVC exposure ($p = 0.011$), but the triple sod-1;sod-4;sod-5 mutant was not ($p = 0.229$) (Fig. 5D). The sod-2 mutant was more sensitive than the sod-2; sod-3 mutant, consistent with a previous report [41]. Based on this data, we used the sod-2 and sod-1;sod-4;sod-5 mutants to examine the role of mitochondrial and cytosolic superoxide dismutase proteins in the response to UVC exposure. We confirmed the growth delay at the molecular level by measuring mitochondrial and nuclear DNA content in wild-type and SOD mutants after UVC exposure. There was a significant delay in mtDNA:nucDNA copy number expansion in the sod-2 worms, but not the sod-1;sod-4;sod-5 worms at 48-h post-exposure (Fig. 5E). The reduction in this ratio was due to the reduction in mtDNA copy number expansion (Fig. S6A), as nucDNA copy number expansion was comparable in all three nematode strains post-UVC exposure (Fig. S6B). mtDNA damage at 24 h post-exposure was significantly increased in the sod-2 mutants compared to N2, but not in sod-1;sod-4;sod-5 mutants (Fig. 5F). Importantly, neither sod-2 mutant or sod-1;sod-4;sod-5 mutant non-exposed controls showed any change in mtDNA damage compared to N2 controls (Fig. S6C). We detected no statistically significant differences between strains in nucDNA damage (Fig. S6D). Given the evidence of increased mitochondrial ROS, we looked at transcript levels of heat shock proteins in the N2, sod-2, and sod-1;sod-4; sod-5 strains to test for evidence of the heat shock response in different subcellular compartments, as well as heat shock contributions to ROS induction [42]. However, we did not detect any sod-2-specific increases in any tested mitochondrial (hsp-6, hsp-60a, and hsp-60b), cytosolic (hsp-16.12 and hsp-16.41), or endoplasmic reticular (hsp-4) heat shock protein (Figs. S6E–J). Thus, this experiment failed to support the hypothesis that an induction of the heat shock response in UVC-treated animals contributed to mitochondrial phenotypes or an increase in ROS production.

2.8. Skn-1 mediates response to UVC exposure

The delayed development and indications of increased mtDNA damage following UVC exposure in the mitochondrial, but not cytosolic SOD mutants could be explained in two ways: either mitochondrial $O_2^{•−}$ causes the growth delay by causing oxidative damage; or conversion of mitochondrial $O_2^{•−}$ to $H_2O_2$ is important in reducing the extent of the growth delay via adaptive signaling. Both $H_2O_2$ and $O_2^{•−}$ can cause macromolecular damage, while only $H_2O_2$ can exit the mitochondria to act as a signaling molecule. To address the first hypothesis, we tested whether antioxidant treatment of 10 mM N-acetyl cysteine [43], 3 mM Trolox [44-46] or 1 and 5 μM MitoQ [47,48] (all previously shown to protect against various stressors in C. elegans) rescued UVC-induced growth delays. We observed no effect of antioxidant treatment on the UVC-induced growth delay (Fig. S7). To test the hypothesis that redox signaling is protective after UVC exposure, we examined the sensitivity of nematodes deficient in skn-1, the nematode homologue of the mammalian oxidant-responsive Nrf2 transcription factor. Nrf2 controls
signaling, more than oxidative damage, is responsible for mediating the persistent phenotypes observed, and found that the 

skn-1 mutants to determine if skn-1 signaling was responsible for the persistent phenotypes observed, and found that the skn-1 mutants were no more sensitive to reduction in ATP levels after UVC exposure than N2 (Fig. 6B), which is also inconsistent with an oxidative damage theory, since skn-1 mutants are oxidative stress-sensitive.

2.9. Early life mtDNA damage results in decreased stress resistance in adulthood

As described earlier, there is evidence for both negative and positive effects of mitochondrial dysfunction and toxicity during development. Developmental mtDNA damage resulted in many of the same phenotypes as mitohormesis, suggesting the potential for a common mechanism. However, we found that although adult size was significantly (10–20%) reduced in response to developmental mtDNA damage (Fig. 6C) and ATP levels were reduced (Fig. 2 and S6), consistent with mitohormetic phenotypes, lifespan was statistically unchanged from that of control nematodes (Fig. 1B), whereas mitohormesis is associated with lifespan extension. Median lifespans were 15 days in both control and UVC-treated animals, and maximal lifespans were 20 and 22 days, in control and UVC-treated animals, respectively. However, organisms outside of the laboratory are exposed to a variety of stressors, and there is evidence for “mitohormetic” responses resulting in increased stress resistance [55]. Instead of increased stress resistance, however, we found that UVC-treated nematodes were more sensitive to rotenone (Fig. 6D) than their untreated counterparts, though we detected no difference in sensitivity to paraquat (Fig. 6E).

3. Discussion

3.1. DOHaD or lifespan extension and stress resistance

Our results indicate that irreparable mtDNA damage incurred early in C. elegans development can lead to a lifelong reduction of mtDNA:
UVC. UVC exposure in early life leads to mtDNA damage that may be sensed by the cell. This damage, or unidentified sensor activity, leads to ROS generated from the ETC. Protected from UVC-induced decrease in ATP levels 4 days post UVC exposure. Two-way ANOVA with Bonferroni correction for multiple comparisons (10 controls. Letters show which groups are significantly different (p ≤ 0.05). Two-way ANOVA with Tukey’s correction for multiple comparisons. n = 60–70 individuals in 2 experimental replicates. D. UVC-exposed nematodes are more sensitive to the complex I inhibitor rotenone, but not to E. paradoxaq. n = 4 per group. Two-way ANOVA with Bonferroni correction for multiple comparisons (13 in D and 10 in E). + p ≤ 0.05 compared to no drug treatment, no UVC exposure control. F. Model for ROS involvement in the development of lifelong effects as a result of early-life exposure to UVC. UVC exposure in early life leads to mtDNA damage that may be sensed by the cell. This damage, or unidentified sensor activity, leads to ROS generated from the ETC. skn-1 is activated by increased ROS generation, leading to decreased ATP and persistent phenotypes in adults.

More broadly, we suggest that observations of “mitohormesis” should be interpreted cautiously. RNAi experiments are different than what is seen in most human mitochondrial diseases, most of which result from mutations in mitochondrial gene, rather than reductions in transcript and protein levels. That mutations and RNAi knockdown are not identical is further underscored by the fact that in C. elegans, many strains that harbor mutations in genes that when knocked down with RNAi result in major losses of protein or function. In our experiments, it appears that transcripts and proteins coded in the mtDNA were produced at sufficient levels despite the presence of transcription-blocking photolesions, presumably due to the multiplicity of mtDNAs. It may be that most environmental chemical exposures belong to a mechanistic category distinct from and possibly poorly modeled by gene knockdown, gene knockout, or specific pharmacological mitochondrial poisoning, because most environmental mitotoxicants are relatively non-specific [8].

3.2. A role for mitochondrial reactive oxygen species and mitochondrial retrograde signaling in metabolic programming

One of the most striking effects observed in response to early life...
mtDNA damage was the persistent reduction in steady state ATP levels, especially since this persisted even after mtDNA lesions were removed, and mtDNA-encoded ETC transcript levels were at or above those in untreated nematodes. mtDNA copy number was reduced at this time, but only approximately 20%, which by itself is not sufficient to reduce ATP levels in *C. elegans* [67]. Similarly, the lack of change in mtRNA levels or activation of UPR [68], suggest that the changes in ATP levels and other phenotypic effects result from signaling, rather than being direct functional deficits resulting from mtDNA damage or copy number changes.

Increased protein oxidation, increased oxygen consumption despite decreased ATP levels and movement, and the reduced sensitivity of *skn-1* mutants to developmental mtDNA damage-induced growth delay, all support mtROS signaling as a mechanism for the regulation of growth delay (Fig. 6F). Increased oxidative stress is also consistent with metabolomic shifts suggesting activation of the malate-aspartate shuttle or increased glycolysis, which could explain some of the increases observed in non-mitochondrial oxygen consumption. In the context of oxidative stress resulting from UVC-induced mtDNA damage, much of the reducing equivalent pool in the mitochondria could be used for reducing oxidized glutathione, making less available for OXPHOS and ATP production, while at the same time reducing the mitochondrial membrane potential or limiting availability of reducing equivalents for the ETC, consistent with a reduced spare respiratory capacity.

Our data suggest that developmental UVC exposure results in a low level of mtROS that plays two roles, both causing damage and acting as signaling molecules, with a dominant long-term role for alterations in developmental ROS signaling. ROS play an important signaling role during development, and disruption of this signaling can alter differentiation and proliferation [68]. Short term exposure to H$_2$O$_2$ in young nematodes results in many similar outcomes to those seen here, including reduced movement and ATP levels, though these effects are transient [69]. We hypothesize that in our experiments, aberrant ROS generation during development led to *skn-1*-mediated metabolic alterations that program the developing nematode for life. In wild-type worms, the toxic effects of O$_2^-$ are ameliorated by the action of SODs, and thus insufficient to cause developmental delay, as indicated both by the sensitivity of the mitochondrial SOD mutants and by the inability of pharmacological antioxidants to rescue growth delay. mtROS, however, appears to be required for a SKN-1-mediated growth delay and metabolic remodeling, as indicated by the initially counterintuitive finding that *skn-1* mutants are protected from developmental delay, and by the absence of a reduction in ATP levels in UVC-exposed *skn-1* mutants. More broadly, given that ROS are both damaging and signaling molecules, it is not surprising that alterations in either direction of their tightly regulated developmental levels may have short-term (i.e., growth) and long-term (i.e., metabolic programming) consequences. Finally, the fact that *afs-1* mutants were also less sensitive to UVC-induced growth delay suggests a role for this transcription factor in mediating growth delay.

3.3. Unexpected mtDNA damage and copy number dynamics

The patterns of change in mtDNA copy number and mtDNA damage were also unexpected in a number of ways. With regards to mtDNA damage in the *glp-1* germine-deficient strains (Fig. 1D), the intriguing reappearance of mtDNA lesions 8 days post-exposure suggests that mtDNA damage may accumulate more quickly with aging in the context of early life exposure than in control animals (in which we did not detect damage at 8 days), possibly due to persistently increased oxidative stress or reduced repair capacity due to reduced ATP or nucleotide availability.

The complete (at least within our limits of detection) removal of photosensitons from mtDNA in developing larvae is in contrast to previous work from our lab in somatic cells of adult nematodes, which had shown just under 40% damage removal after 72 h in adult nematodes [13]. mtDNA replication appears to stop or at least fall below detection in adult nematodes, and copy number slowly declines with a half-life of approximately 10 days [70]. This half-life roughly corresponds to the rate at which damage is removed, suggesting that mtDNA damage removal in adult nematodes is tied to turnover of mitochondrial genomes. In contrast, mtDNA replication is essential and extensive during development. We hypothesize that increased mtDNA replication (leading to dilution of damaged genomes with new, undamaged copies) and, potentially, more frequent turnover leads to more complete reduction in per-nucleotide levels of damage during development. Supporting this hypothesis, the increase in mtDNA copy number between days 0 and 2 (3.88-fold) is proportionally similar to the reduction in lesion frequency over the same time period (3.98-fold). This would suggest that photodimers in mtDNA are not removed in any quantitatively important fashion during the first 2 days of *C. elegans* development.

With regard to mtDNA copy number, complete lesion removal and induction of polg-1 suggest that mtDNA replication should be unimpeded, yet reductions in mtDNA:nuDNA copy number persist throughout adulthood. However, mtDNA replication appears to cease between 3 and 6 days of age in untreated nematodes [70]. Therefore, it is possible that replication of some mtDNAs may be blocked by lesions until the point at which replication stops, resulting in fewer mtDNA copies per cell. Alternately, it is possible that a reduction in mtDNA copy number results from increased removal. Consistent with this theory, oxidative damage in cultured mammalian cells induced strand breaks in mtDNA, leading to degradation of mitochondrial genomes [71].

A limitation of our study is the inability to draw conclusions about cell-type specific effects of either mtDNA damage or copy number which may be important because mitochondria vary significantly between cell types [72]. Cell- and tissue-specific differences include turnover rates [73] and levels of mitophagy [74], such that damaged genomes or alterations in copy number may persist only in some cell types. This is an intriguing hypothesis, as mitochondrial dysfunction specifically in neurons results in whole organism effects in *C. elegans* [56], and is implicated in many human diseases [75].

4. Conclusions

Overall, these results identify life-long redox signaling-mediated alterations in mitochondrial function resulting from a low-level exposure that had no effect on lifespan. The outcomes later in life were largely deleterious, although we cannot exclude that the metabolic changes we observed may be advantageous under certain circumstances. These results are potentially important in the context of human health, because environmental exposures that result in irreparable mtDNA damage are common [12], there are increasing reports of pollutants affecting mitochondria [8,76], and we now understand that pollution is the major environmental driver of loss of life globally [5], despite the fact that we know very little about the toxic effects of the great majority of anthropogenically-produced chemicals [7].

5. Methods

5.1. *C. elegans* strains and culture conditions

Populations of *C. elegans* were maintained on K-agar plates seeded with *E. coli* OP50 bacteria, unless otherwise noted. N2 (wild-type), JK1107 *glp-1(q224)*, *sod-2* (gk257 I), *sod-3* (gk235), gas-1 (f21), mev-1 (Kn1), *glp-1* (e2141), daf-16 (mbu6), *skn-1* (zu670), SJ4100 zclS13 (*hsp-6;GFP*), SJ4058 zclS9 (*hsp-60;GFP*), and pkc-1 (gk1) were obtained from the Caenorhabditis Genetics Center (CGC), University of Minnesota. PE255 *glp-4(bn2)* were provided by Christina Lagido, University of Aberdeen (Aberdeen, UK). The sod-2;sod-3 (sod-2 (gk257 I); sod-3 (tm760 X) double mutant strain and sod-1;sod-4;sod-5 triple mutant strain (sod-1 (tm776);sod-5 (tm1146) II; sod-4 (gk101) III) were a kind gift from Bart Braeckman (Ghent University, Ghent, Belgium).
5.2. UVC exposure, DNA damage and genome copy number

UVC exposures were conducted in a custom-built exposure cabinet as previously described [13]. UVC output was measured with a UVC radiometer (UVP, Upland, CA). Synchronized (by bleach/NaOH egg isolation [77]) L1 nematodes were maintained on plates that contain no bacterial food, and therefore do not develop for UVC exposures, and were then transferred to seeded plates, as previously described [14]. Mitochondrial and nuclear DNA damage levels and mtDNA copy number were measured as previously described [78]. All PCR conditions and primer sequences can be found in Table S1.

5.3. ATP levels and oxygen consumption

ATP levels were measured in two different strains by two methods. First, using the JK1107 glp-1(q224) strain, ATP levels were measured as described in [79] using the Molecular Probes ATP determination Kit (Invitrogen/Life Technologies, Carlsbad, CA, USA). Second, the firefly luciferase expressing PE255 glp-4(bn2) strain was used to investigate relative, steady state ATP levels in vivo, in live nematodes at 2, 4, 6, 8, 10 and 12 days post final UVC dose, as previously described [80]. Details of both protocols are available in the supplemental methods section. Oxygen consumption was measured using a Seahorse Biosciences XF24 extracellular flux analyzer as described [81].

5.4. Gene expression assays

RNA was isolated from between 1000 and 2000 nematodes according to the Qiagen RNeasy Min Kit protocol. cDNA was created from 100 ng of the isolated RNA using the High Capacity Reverse Transcription kit (Life Technologies) per the manufacturer’s instructions. Gene expression was measured via real-time PCR using the Power SYBR Green PCR Master Mix (Life Technologies). Changes in expression levels were calculated based on the standard delta-delta-CT method, compared to housekeeping genes cdc-42 and pmp-3. Primer sequences and PCR conditions can be found in Table S2.

5.5. Targeted metabolomics

Nematodes (approximately 7500 for each sample) were washed, incubated in K-medium at 20 °C for 30 min to allow for gut clearance, and washed twice in ice-cold PBS. Worms were resuspended in 0.6% formic acid and stored at –80 °C. Samples were thawed on ice, lyed by sonication and aliquots were removed for total protein determination. 270 µL acetonitrile was added to each sample, and they were vortexed for 1 min and centrifuged at 15,000×g for 10 min to pellet proteins. Amino acids, acylcarnitines and organic acids were analyzed using stable isotope dilution technique. Amino acids and acylcarnitine measurements were made by flow injection tandem mass spectrometry using sample preparation methods described previously [82,83]. The data were acquired using a Waters TQD mass spectrometer equipped with AcquityTM UPLC system and controlled by Masslynx 4.1 operating system (Waters, Milford, MA). Organic acids were quantified using methods described previously [84] employing Trace Ultra GC coupled to ISQ MS operating under Xcalibur 2.2 (Thermo Fisher Scientific, Austin, TX). Metabolite levels were normalized to total protein as determined by BCA (Thermo Scientific, Rockford, IL) and fold changes compared to day 4 control samples were calculated.

5.6. Mitochondrial isolation and oxyblot

Mitochondria were isolated essentially as described [85]. Briefly, worms were washed off plates, rinsed twice in K-media and allowed to clear guts for 30 min in K-medium at room temperature. All steps from here were conducted on ice. Worms were washed once in K-medium and twice in 10 mL MSM-E buffer (MSM-E – 150 mL MSM (=40.08 g mannitol, 23.96 g sucrose, 1.047 g MOPS, 1L milliQ H2O, pH 7.4) plus 3 mL 0.1 M EDTA), resuspended in 1 mL MSM-E and lysed in a glass-Teflon potter homogenizer. Worms were microscopically monitored for lysis. Once sufficiently lysed, one volume MSM-EB (50 mL MSM-E + 0.2 g fatty acid-free BSA) was added and samples were centrifuged at 300 × g for 10 min. Supernatant was transferred to a fresh tube and kept, and pellet was re-extracted. Supernatants from the two extractions were combined and spun at 7000 × g for 10 min to yield the mitochondrial pellet, and supernatant samples were taken as the cytoplasmic fraction. Mitochondrial pellets were washed once in MSM-E and once in MSM. Oxyblot was performed following the manufacturer’s instructions (MilliporeSigma, Burlington, MA).

5.7. Growth, antioxidant rescue, protein translation exacerbation, and lethality assays

Laval growth was assessed by measuring nematode size at 72 h post-exposure, using either a COPAS Biosorter (Union Biometrica, Holliston, MA), which measures extinction and time of flight of individual C. elegans [46], or microscopic imaging. For the latter, aliquots of nematodes were frozen, thawed and imaged at 10× magnification on a Zeiss Axioskop. Nematode length was measured using NIS elements BR software (Nikon Inc. Melville, NY, USA). In many cases, growth is presented as percent control to permit comparisons between the methods and between strains that may develop at different rates under control conditions. In some experiments, larvae were staged (L1 through adult) at 24, 48, 72, and 96 h post-final exposure. For chemical exacerbation experiments, the mitochondrial translation inhibitors doxycycline and chloramphenicol were added after plates cooled, and nematodes were exposed to chemicals throughout development after UVC exposure. For trolox, MitOQ, and N-acetyl cysteine antioxidant rescue experiments, compounds were added to the medium prior to pouring plates, nematodes were exposed to antioxidants throughout development after UVC exposure, and size was measured on the COPAS Biosorter. Lethality assays were conducted 4 days after the 3rd UVC exposure. 10–15 nematodes were transferred to plates containing either rotenone or paraquat that were seeded with UVC-inactivated UvrA deficient OP50 E. coli to reduce bacterial metabolism of the chemicals [86]. Nematodes were scored for survival 24 h later and were counted as dead if they did not respond (move) to gentle prodding with a worm pick.

5.8. Lifespan analysis

Lifespan was determined on K agar plates with an OP50 lawn at 20 °C. Approximately 25 synchronized (as above) L1 nematodes were plated per condition were assayed in triplicate experiments. Beginning one day after reaching L4 and continuing until reproduction stopped, all adults were transferred daily onto a new plate, leaving offspring on the old plate. Nematodes were monitored daily by tapping the adults on the head. Animals were considered dead if no movement was observed following repeated probing. Individual lifespans were calculated from egg (day 0) until death.

5.9. Analysis of movement

Synchronized glp-1 L1 larvae were obtained by bleach-sodium hydroxide isolation of eggs (described above), transferred to unseeded n-peptone K-agar, and exposed to 3 sequential doses, 24 h apart, of 0, 3, or 4 J/m² UVC as described above. Next, nematodes were transferred to K-agar plates seeded with OP50 and were grown at 25 °C for either 4 or 8 days. For each treatment group, six 60-s videos of nematode locomotion were recorded on a Nikon SMZ1500 stereomicroscope using NIS-Elements BR software. Analysis was completed using Fiji/ImageJ. Nematode pathways were constructed using the first 10 s of each video and quantified using the Anlyze Particles function. Pathway area was normalized to mean nematode size for each treatment group.
5.10. Measurement of mitochondrial unfolded protein response

UPR
\textsuperscript{mt} activation was measured in hsp-6 (strain SJ4100) and hsp-60 (strain SJ4058) GFP fusion reporter strain 4 days post 2.5 or 5 J/m\textsuperscript{2} UVC exposure. Chloramphenicol exposure was included as a positive control. Worms were washed with K-medium, and approximately 100 worms in 100 μL K-medium were aliquoted into wells of a white 96 well plate (70 worms per well for SW4100 chloramphenicol exposure). GFP fluorescence was measured from 4 wells on a FLUOstar Optima microplate reader and adjusted for number of worms per well.

5.11. Statistical analysis

Specific analyses are indicated in the manuscript text of figure legends. In Figures, error bars indicate standard errors of the mean. In any case where asterisks are employed, p ≤ 0.05. Correction for multiple comparisons is indicated in the figure legends.

Author contributions

Kathleen A. Hershberger: formal analysis, writing - original draft preparation, writing - review and editing, visualization. John P. Rooney: conceptualization, methodology, formal analysis, investigation, writing - original draft preparation. Elena A. Turner: investigation. Lauren J. Donoghue: investigation. Rakesh Bodhicharla: investigation. Laura L. Maurer: investigation. Ian T. Ryde: investigation. Jina J Kim: investigation. Rashmi Joglekar: investigation. Jonathan D. Hbsman: investigation, writing - review and editing. Latasha L. Smith: investigation. Dbbal P. Bhatt: investigation, Olga R. Ilkayeva: methodology, investigation, resources. Matthew D. Hirschey: supervision. Joel N. Meyer: supervision. Dhaval P. Bhatt: investigation. Olga R. Ilkayeva: methodology, investigation, resources. Matthew D. Hirschey: supervision. Joel N. Meyer: conceptualization, writing - review and editing, supervision, project administration, funding acquisition.

Declaration of competing interest

The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

Acknowledgements

This research was supported by the National Institute of Environmental Health Sciences (R01ES017540, T32ES021432, and P42ES010356). The content is solely the responsibility of the authors and does not necessarily represent the official views of the NIH. We thank Tracey Crocker for assistance with the docycline and chloramphenicol experiments.

Appendix A. Supplementary data

Supplementary data to this article can be found online at https://doi.org/10.1016/j.redox.2021.102000.

References

[1] P.F. Chimney, Mitochondrial disease in adults: what's old and what's new? EMBO Mol. Med. 7 (12) (2015) 1503–1512, https://doi.org/10.15252/emmm.201505079.
[2] A. Suomalainen, P. Isohanni, Mitochondrial DNA deletion syndromes-many genes, common mechanisms, Neuro Muscul. Disord.: NMD 20 (7) (2010) 429–437, https://doi.org/10.1016/j.nmd.2010.03.017.
[3] J. Nunnari, A. Suomalainen, Mitochondria: in sickness and in health, Cell 148 (6) (2012) 1145–1159, https://doi.org/10.1016/j.cell.2012.02.035.
[4] J.A. Dykens, Y. Will, The significance of mitochondrial toxicity testing in drug development, Drug Discov. Today 17 (12–18) (2007) 777–785, https://doi.org/10.1016/j.drudis.2007.07.013.
[5] P.J. Landrigan, R. Fuller, N.J.R. Acosta, O. Adeyi, R. Arnold, N.N. Basu, M. Zhong, The Lancet Commission on pollution and health, Lancet 391 (10119) (2018) 462–512, https://doi.org/10.1016/S0140-6736(17)35245-0.
[6] J.M. Braun, K. Gray, Challenges to studying the health effects of early life environmental chemical exposures on children’s health, PLoS Biol. 15 (12) (2017), https://doi.org/10.1371/journal.pbio.2004810.
[7] L. Gross, L.S. Birnbaum, Regulating toxic chemicals for public and environmental health, PLoS Biol. 15 (12) (2017), e2004814, https://doi.org/10.1371/journal.pbio.2004814.
[8] J.N. Meyer, J.H. Hartman, D.F. Mello, Mitochondrial toxicity, Toxicol. Sci. 162 (1) (2018) 15–23, https://doi.org/10.1093/toxsci/kyx008.
[9] D.A. Roubicek, N.C. de Souza-Pinto, Mitochondria and mitochondrial DNA as relevant targets for environmental contaminants, Toxicology 391 (2017) 100–108, https://doi.org/10.1016/j.tox.2017.06.012.
[10] M.A. Graziewicz, J.M. Sayer, D.M. Jerina, W.C. Copeland, L.K. medium were aliquoted into wells of a white 96 well plate (70 worms per well for SW4100 chloramphenicol exposure). GFP fluorescence was measured from 4 wells on a FLUOstar Optima microplate reader and adjusted for number of worms per well.

11
C.T. Murphy, S.A. McCarroll, C.I. Bargmann, A. Fraser, R.S. Kamath, J. Ahringer, J.D. Hayes, A.T. Dinkova-Kostova, The Nrf2 regulatory network provides an interface between redox and intermediary metabolism, Trends Biochem. Sci. 39 (4) (2014) 133–139, https://doi.org/10.1016/j.tibs.2014.02.008.

J.A. Lewis, J.T. Fleming, Basic culture methods, in: H.F. Epstein, D.C. Shakes (Eds.), Caenorhabditis elegans: Modern Biological Analysis of an Organism, Cold Spring Harbor Laboratory Press, Cold Spring Harbor, NY, 1995, pp. 3–29.

J.P. Rooney, I.T. Ryde, C. Anbalagan, R. Joglekar, J.N. Meyer, PCR-based analysis of mitochondrial DNA copy number, mitochondrial DNA – Mitochondrion 10 (2) (2009) 154–167, https://doi.org/10.1089/mito.2008.0531.

K. Brys, N. Castelein, F. Matthijssens, J.R. Vanfleteren, B.P. Braeckman, Disruption of the mitochondrial DNA copy number in C. elegans, Exp. Gerontol. 56 (2014) 69–76, https://doi.org/10.1016/j.exger.2014.10.008.

J.A. Lewis, J.T. Fleming, Basic culture methods, in: H.F. Epstein, D.C. Shakes (Eds.), Caenorhabditis elegans: Modern Biological Analysis of an Organism, Cold Spring Harbor Laboratory Press, Cold Spring Harbor, NY, 1995, pp. 3–29.
[80] C. Lagido, J. Pettitt, A. Flett, L.A. Glover, Bridging the phenotypic gap: real-time assessment of mitochondrial function and metabolism of the nematode Caenorhabditis elegans, BMC Physiol. 8 (2008) 7. Retrieved from, http://www.ncbi.nlm.nih.gov/entrez/query.fcgi?cmd=Retrieve&db=PubMed&dopt=Citation &list_uids=18384668.

[81] A.L. Luz, J.P. Rooney, L.L. Kubik, C.P. Gonzalez, D.H. Song, J.N. Meyer, Mitochondrial morphology and fundamental parameters of the mitochondrial respiratory chain are altered in Caenorhabditis elegans strains deficient in mitochondrial dynamics and homeostasis processes, PLoS One 10 (6) (2015), e0130940, https://doi.org/10.1371/journal.pone.0130940.

[82] J. An, D.M. Muoio, M. Shiota, Y. Fujimoto, G.W. Cline, G.I. Shulman, C.B. Newgard, Hepatic expression of malonyl-CoA decarboxylase reverses muscle, liver and whole-animal insulin resistance, Nat. Med. 10 (3) (2004) 268–274, https://doi.org/10.1038/nm995.

[83] C.T. Ferrara, P. Wang, E.C. Neto, R.D. Stevens, J.R. Bain, B.R. Wenner, A.D. Attie, Genetic networks of liver metabolism revealed by integration of metabolic and transcriptional profiling, PLoS Genet. 4 (3) (2008), e1000034, https://doi.org/10.1371/journal.pgen.1000034.

[84] M.V. Jensen, J.W. Joseph, O. Ilkayeva, S. Burgess, D. Lu, S.M. Ronnebaum, C.B. Newgard, Compensatory responses to pyruvate carboxylase suppression in islet beta-cells. Preservation of glucose-stimulated insulin secretion, J. Biol. Chem. 281 (31) (2006) 22342–22351, https://doi.org/10.1074/jbc.M604350200.

[85] M.J. Falk, E.B. Kayser, P.G. Morgan, M.M. Sedensky, Mitochondrial complex I function modulates volatile anesthetic sensitivity in C. elegans, Curr. Biol. 16 (16) (2006) 1641–1645, https://doi.org/10.1016/j.cub.2006.06.072.

[86] L.L. Maurer, I.T. Ryde, X. Yang, J.N. Meyer, Caenorhabditis elegans as a model for toxic effects of nanoparticles: lethality, growth, and reproduction, Curr Protoc Toxicol 66 (2015), https://doi.org/10.1002/0471148856.ct2010666, 2010 21-20 10 25.