The helminth fauna of brown trout (*Salmo trutta*) from a sub-alpine lake revisited after 40 years with introduced European minnow (*Phoxinus phoxinus*)

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Borgstrøm R, Hatleli Mestrand Ø, Brittain JE and Lien L. 2021. The helminth fauna of brown trout (*Salmo trutta*) from a sub-alpine lake revisited after 40 years with introduced European minnow (*Phoxinus phoxinus*). Fauna norvegica 41: 15–26.

The helminth fauna of brown trout (*Salmo trutta*) in the Norwegian subalpine lake, Øvre Heimdalsvatn was studied by examination of gills, eyes, body cavity, kidney, stomach, pyloric region and intestine in a total of 112 brown trout randomly sampled in June, July, and September 2011. Ten helminth species, *Discocotyle sagittata*, *Phyllobothrium umblicae*, *Crepidostomum farionis*, *C. metoecus*, *Diplostomum sp*., *Proteocephalus sp*., *Cystodistomum truncatus*, *Dibothriocotyle ditemus*, *D. denticicus*, and *Capillaria sp*. were identified. These data were compared to data from the period 1969 to 1972, just after the first record of the European minnow (*Phoxinus phoxinus*) in 1969. All ten helminth species, except *D. denticicus*, were also present in 1969–72. However, a few major changes in infection intensities have occurred. The cestode *D. ditemus* and the trematode *Diplostomum sp*., both with piscivorous birds as final hosts, had markedly higher relative densities (abundance) in brown trout in 2011 compared to 1969–72, while the two *Crepidostomum* species showed a substantial decline in relative densities. We suggest that these changes may be indirectly related to the establishment and subsequent population increase of European minnow in the lake. The abundance of minnows may have increased the food basis for the piscivorous birds, primarily mergansers and the black-throated diver that now regularly forage in the lake. In addition, there have been changes in the littoral invertebrate community, including species serving as intermediate hosts of some of the brown trout parasites.

doi: 10.5324/fn.v41i0.3952. Received: 2021-01-21. Accepted: 2021-03-23. Published online: 2021-05-27.

ISSN: 1891-5396 (electronic).

Keywords: parasite fauna, parasite abundance, fish, long-term changes, invasive species, mountain lake

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INTRODUCTION

Fish may serve as both intermediate and final hosts of helminth species which have developed a variety of life history strategies, from complex life cycles with several intermediate hosts to direct cycles without intermediate hosts (Moravec 1994; Margolis 1995; Hoffman 1999; Amundsen et al. 2003). Accordingly, environmental changes as well as changes in abundance of both intermediate and final hosts may all influence the parasite burden in a fish population.

During the last hundred years there has been a worldwide spread of exotic freshwater fish and invertebrate species (Esch et al. 1988; Welcomme 1988; Hall & Mills 2000). Such introductions may result in substantial changes in the parasite fauna and infection patterns of indigenous fish species (Gozlan et al. 2005). In Norway, the European minnow (*Phoxinus phoxinus* (Linnaeus, 1758)) has been spread by man to localities far outside its natural distribution area (Saltveit & Brabrand 1991; Hesthagen & Sandlund 1997). At present, this small cyprinid has become established in all Norwegian counties, with populations in several mountain lakes where brown trout (*Salmo trutta* Linnaeus, 1758) was previously the sole fish species (Museth et al. 2007). The benthic invertebrate community may be strongly influenced by the establishment of a European minnow population, as observed in the Norwegian subalpine lake, Øvre Heimdalsvatn (Næstad & Brittain 2010). Despite extensive fishery studies in this lake since 1957 (Jensen 1977), the European minnow was only first recorded in 1969 (Larsson et al. 1978), and in the following years established a large population (Lien 1981; Museth et al. 2002). Young brown trout and European minnow share both the streams and the shallow littoral areas of the lake, and the overlapping habitat use and diet of these two species suggest a negative influence of the minnows on the trout population (Borgstrøm et al. 2010; Museth et al. 2010; Næstad & Brittain 2010).

The summer diet of brown trout during the years 1969–1972 included a high proportion of *Gammarus lacustris* G. O. Sars, 1863 and *Lepidurus arcticus* (Pallas, 1793), while fish were never recorded
as part of the diet (Lien 1978). However, *L. arcticus* has been almost absent in brown trout diet during the last decades (Borgstrøm et al. 2010), and *G. lacustris* has declined considerably in abundance in the shallow part of the littoral zone (Næstad & Brittain 2010). Nevertheless, *G. lacustris* is still a common food item of brown trout (Borgstrøm et al. 2010). Minnows have become an important part of the trout diet, but also small brown trout are eaten by larger conspecifics (Borgstrøm et al. 2010).

The parasites of brown trout in the lake were first studied by monthly samples during the period December 1969 to December 1971, with some additional samples in 1972 (Lien 1976; 1978), i.e. during a period when European minnow still had a very low population density (Lien 1981). These data form a basis for evaluating possible long-term changes in the parasite fauna of brown trout in the lake, from 1969–72 to 2011. Brown trout serve as intermediate host for some of the helminth species recorded in 1969–72, with piscivorous birds as final hosts. Water temperature may be an important environmental factor, influencing development of parasite stages (Karvonen et al. 2010, 2013), and thereby affecting both prevalence and intensity of infection in brown trout. Thus, we hypothesise that environmental changes and changes in the invertebrate community, together with changes in dynamics of the fish populations and abundance of piscivorous birds, have influenced the helminth fauna of brown trout in the lake, Øvre Heimdalsvatn.

**MATERIAL AND METHODS**

**The lake Øvre Heimdalsvatn**

The subalpine lake, Øvre Heimdalsvatn is situated in southern Norway (latitude 61°25′32″ and longitude 8°52′10″), at 1088 m a. s. l., on the eastern slopes of the Jotunheimen Mountains (Figure 1). The lake is normally ice-covered from the end of October until early June (Kvambekk & Melvold 2010). The surface area is 0.78 km², with a maximum depth 13.0 m, and an average depth 4.7 m (Grøterud & Kloster 1978). Lake temperatures in Øvre Heimdalsvatn were measured continuously in 1970–72 (Kloster 1978), and from 1986 temperature in the outlet river (Hinøgla) has been measured daily at 11 a.m. (Sildre database; [www.nve.no](http://www.nve.no)), but not in the lake itself. According to Kvambekk & Melvold (2010) there is a significant correlation between air temperature at Skåbu meteorological station ([www.seklima.met.no](http://www.seklima.met.no)) and the outlet temperature of Øvre Heimdalsvatn during the months July - September. Data from Skåbu are available from 1969 to the end of 2009 (the station is now closed). The average June - August temperature at Skåbu shows a weak declining trend from 1969 to 2009, but with large annual variations, with 1997 as the warmest and 1993 as the coldest summers during this period (Figure 2). The average temperature for the summers (June - August) 1969–72 and 2006–2009 was 10.8 and 10.7°C, respectively.

Øvre Heimdalsvatn is a reference lake for subalpine ecosystems, with long time series of physical, chemical and biological parameters (Brittain & Borgstrøm 2010a, 2015; Brittain et al. 2019). The brown trout population studies commenced in 1957 (Jensen 1977), and at intervals they have continued until today (summarised in Brittain & Borgstrøm 2015; Brittain et al. 2019). In 1957, the brown trout population was very dense (>20 kg/ha), with high annual recruitment and early growth stagnation of individual fish. After extensive exploitation during the following years, few fish older than ten years were present, and with a marked increase in individual growth rates (Jensen 1977). Annual recruitment to the brown trout population has

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**Figure 1.** The lake, Øvre Heimdalsvatn, 1088 m a. s. l., still partly ice-covered on 12 June 2008. Some of the peaks of the Jotunheimen Mountains in the background.
been halved during the years 1993–2009 compared to the years 1963–1970 (Jensen 1977; Borgstrøm et al. 2010), but exploitation has been negligible, and total biomass has been in the same order compared to the years 1963–1970 (Jensen 1977; Brittain et al. 2019).

The benthic fauna and diet of brown trout was first investigated by the freshwater section of the Norwegian IBP (International Biological Programme) during the period 1969–72 (Vik 1978) and has been continued at irregular intervals since (see Brittain & Borgstrøm 2010b; Brittain et al. 2019).

The number of fish-eating birds at the lake has also changed from the initial period, coconmitant with the establishment of European minnow. In 1968–1972, the Common gull (Larus canus Linnaeus, 1758) together with the mergansers, Mergus merganser Linnaeus, 1758 and M. serratator Linnaeus, 1958, were observed, but they were not common (Lien & Nydal 1973). During the 1990s and up until the present study in 2011, three to five females of M. merganser and M. serratator have been feeding and nesting in close proximity to the lake (Jan T. Lifjeld pers. comm., and own observations). The black-throated diver (Gavia arctica Linnaeus, 1758) has also been regularly foraging there in these years (Jan T. Lifjeld, pers. comm., and own observations), while only two observations were made during the years 1968–72 (Lien & Nydal 1973). Likewise, grey heron (Ardea cinerea Linnaeus, 1758) has been observed regularly (own observations) but was not recorded by Lien & Nydal (1973). Observations of Arctic tern (Sterna paradisaea Pontoppidan, 1793) have increased, with one to three individuals foraging in the lake, in addition to about five nesting pairs of the Common gull (Jan T. Lifjeld pers. comm.).

A total number of 330 brown trout captured by gillnetting with various mesh sizes all over the lake during the years 1969–72 were quantitatively examined for helminth parasites by Lien (1976, 1978), as part of the IBP on Øvre Heimdalsvatn. Ten trout were examined each month during the period December 1969 to December 1971, and 25 trout were examined in each of the months April, August, and December 1972. In this extensive sampling of parasites from brown trout, 10 helminth species were recorded (Lien 1976, 1978), of which brown trout is the final host of six species, and piscivorous birds of four species (Table 1). Invertebrates are included in the life history of all the species (Table 1), except Discocotyle sagittata (Leuckart, 1842) which has a direct life cycle. Results of the study by Lien (1976) have formed the basis for a comparison with results obtained in the present study in 2011.

**Table 1. Helminth species/genera described from brown trout in the lake Øvre Heimdalsvatn (Lien 1976, 1978), with an overview of species involved in their assumed life histories in the lake.**

| Species/genus                     | Final host/organ | First intermediate host | Second intermediate host |
|----------------------------------|------------------|-------------------------|--------------------------|
| Discocotyle sagittata (Leuckart 1842) | Brown trout, gills | Bivalves (Sphaerium Scopoli, 1777) and Psidium Pfeiffer, 1821 |                          |
| Diesing, 1850                    |                  |                         |                          |
| Phyllobothrium umbilae (=conostomum) (Fabricius, 1780) | Brown trout, kidney | Bivalves (Psidium spp., and snail, Radix balthica (Linnaeus, 1758)) | Mayflies (Ephemeroptera), Gammarus lacustris |
| Crepidostomum metoecus (Braun, 1900) and C. farionis (Müller, 1780) Lühe, 1900 | Brown trout, intestine | Snail, Radix balthica (Linnaeus, 1758) | Brown trout, European minnow, in eyes |
| Diplostomum von Nordmann, 1832 | Piscivorous birds, intestine | |                          |
| Proteinophasus Weinland, 1858 | Brown trout, intestine/ pyloric caeca | Copepods (Cyclops Müller, 1776) |                          |
| Dibothriophalus ditremus (Creplin, 1825) Lühe, 1899, and D. dendriticus (Nitzsch, 1824) Lühe, 1899 | Piscivorous birds, intestine | Copepods (Cyclops spp.) | Brown trout, organs in body cavity |
| Cystocephalus truncatus (Pallas, 1781) | Brown trout, pyloric caeca | Gammarus lacustris |                          |
| Capillaria Zeder, 1800 | Brown trout, intestine | Probably oligochaetes |                          |
| Eustrongylides Jägerskiöld, 1909 | Piscivorous birds, intestine | Oligochaetes | Brown trout, European minnow, in body cavity |

**Figure 2.** Mean air temperature for the months June - August at Skåbu meteorological station 1969–2009 (data from https://seklima.met.no).
A total of 392 brown trout were caught, of which 112 were randomly picked for sampling of helminths from the organs, gills, eyes, kidney, stomach, pyloric region, intestine, and body cavity of the sampled fish. For collecting *Proteocephalus* Weinland, 1858 and *Cystocephalus truncatus* (Pallas, 1781) which both have their scolex attached in the pylorus caeca, the pylorus region was opened, loose contents removed for examination of helminth individuals, and subsequently this part of the intestine was placed in water overnight to release individuals of the two cestodes. Examination and searching of the organs for helminth parasites is time consuming, and we had to take subsamples for specific species/organs even among the 112 individuals. Before a fish was selected for parasite examination, it was decided what organs that should be included. Based on the age-length relationship of brown trout from the lake, the examined fish were collected to ensure sampling from a range of age and length classes. The sampled helminths from each examined brown trout were counted and preserved in ethanol. In total, brown trout with age from 3 to 15 winters are included in the study (Table 2). Due to the stagnation in annual growth in length after maturation, the differences in mean length of the age-classes ≥ 8 winters are small (Table 2), as also shown in Solhaug et al. (2010). Therefore, we sampled a relatively high number of fish in the length range 30.0–39.9 cm to ensure that several age-classes were represented with at least some individuals. Lien (1976) presented the relative density (abundance) of parasite species recorded (wrongly stated as infection intensities) in the age-classes 3–4, 5, 6, and ≥ 7 winter old brown trout from the lake Øvre Heimdalsvatn, and to make a comparison with his results, we have used the same age categories in the summary statistics.

The trout stomach and oesophagus contents of the individuals examined for parasites were preserved in ethanol for subsequent dietary analysis in the laboratory. The volume percent of the different prey item categories in each stomach was recorded according to Hynes (1950).

### Parasite identification

The parasites sampled in 2011 were identified to species or genus, and counted under a stereomicroscope, used both in the field laboratory and in the laboratory of the Natural History Museum, University of Oslo. *Crepidostomum farionis* (Müller, 1784) and *C. metoecus* (Braun, 1900) were identified and separated by morphology, according to Moravec (2002), and *Phyllodistomum umblae* (Fabricius, 1780) was identified by use of description given by Bakke (1984). This species (*P. conostomum = P. umblae*) was also identified previously by Bakke & Lien (1978) from brown trout in the lake, Øvre Heimdalsvatn. The plerocercoids of *Dibothriocephalus* Lühe, 1899 (formerly *Diphyllobothrium* Cobbold, 1858, see Waeschenbach et al. 2017) were identified to *D. ditremus* (Creplin, 1825) Lühe, 1899, and *D. dendriticus* (Nitzsch, 1824) Lühe, 1899 using descriptions in Vik (1964), Halvorsen (1970), and Andersen and Gibson (1989), while the other helminth species were identified either to genus or species using Bykhovskaya-Pavlovskaya et al. (1964). Ten helminth species were found in the samples in June, July and September 2011 (Table 3), and of those, nine were also present in 1969–1972 (Table 1). The most numerous parasite taxon in 2011 was *Diplostomum* sp., followed by *Crepidostomum metoecus* (Table 3).

### Statistical analysis

The parasite counts were recorded in Excel spreadsheets and used to compute mean and median intensity, prevalence, and relative density (or abundance, i.e., mean number of individuals of a particular parasite species per host examined), all according to Margolis et al. (1982). In addition, linear regression of relative density of each parasite species in age and length-classes of brown trout were computed.

### RESULTS

#### Intermediate hosts in brown trout diet

The brown trout diet in 2011 varied both seasonally and with fish size (Figure 3). Bivalves (*Pisidium* sp.) i.e., first intermediate host in the life history of *P. umblae* and *Crepidostomum* spp., had a low occurrence in diet of brown trout, and made up a small volume percent of the stomach contents. Likewise, mayflies contributed a low volume percent in the diet, except in the largest length-class in July. *Radix balthica* (Linnaeus, 1758), was relatively common in the diet, although contributing a low volume percent of the stomach contents. *G. lacustris*, the intermediate host of *C. truncatus*, and possibly an intermediate host of *Crepidostomum* spp., made up a relatively large volume percent in the length-classes 20–29.9 and 30–39.9 cm. Copepods, serving as intermediate hosts for *Proteocephalus* sp. and *Dibothriocephalus* sp., were found in one stomach only, in September. European minnow had been eaten by a few trout in the length-classes > 20 cm (Figure 3).

Prey groups not functioning as intermediate hosts for the brown trout helminths, such as chironomids, other aquatic insects, and terrestrial insects, constituted a large part of the brown trout summer diet, together with cladocerans. *Lepidurus arcticus* was present but was found only in a single brown trout in July.

#### The helminth parasite fauna

Ten helminth species were found in the samples in June, July and September 2011 (Table 3), and of those, nine were also present in 1969–1972 (Table 1). The most numerous parasite taxon in 2011 was *Diplostomum* sp., followed by *Crepidostomum metoecus* (Table 3).

**Dibothriocephalus** spp.

The prevalence of *D. ditremus* varied from a minimum of 22.2 % in age-class 5 to a maximum of 60.9 % in age-class 7–8 winters, and except for age-class 3–4, the mean and median infection intensities were relatively high in all age-classes (Table 3). The highest numbers of larvae were found in two 15 winter-old fish, with respectively 89 and 297 plerocercoids. Despite the over-dispersion of larvae, the relative density increased significantly with age (linear regression: $R^2=0.76$, df=6, $F=16.02$, p<0.05), while the linear regression between relative density and length-class was not significant ($R^2=0.52$, df=5, $F=4.40$, p>0.05), due to the very much higher relative density in length-class 35–39.9 cm (Figure 4). A single specimen of *D. dendriticus* was identified as a new species in the lake (Table 3).

### Table 1. Number and average length (cm) of brown trout in age-class 3–15 winters from the lake Øvre Heimdalsvatn sampled for parasites and stomach contents, in June, July, and September 2011.

| Age-class (winters) | Number examined | Mean length (cm) |
|--------------------|-----------------|------------------|
| 3                  | 10              | 15.1             |
| 4                  | 12              | 15.0             |
| 5                  | 16              | 20.8             |
| 6                  | 15              | 23.1             |
| 7                  | 8               | 27.6             |
| 8                  | 15              | 31.0             |
| 9                  | 13              | 32.3             |
| 10                 | 5               | 32.3             |
| 11                 | 8               | 32.8             |
| 12                 | 4               | 33.8             |
| 13                 | 1               | 34.1             |
| 14                 | 3               | 35.0             |
| 15                 | 2               | 38.7             |
Table 3. Mean (Mea.) and median (Med.) intensity, and prevalence (Prev.) of helminths sampled from different age-classes of brown trout in the lake Øvre Heimdalsvatn, as an average for the months June, July, and September 2011.

| Age-class (winters) | 3–4 | 5 | 6 | 7–8 | 9–10 | 11–15 |
|---------------------|-----|---|---|-----|------|-------|
| Mea. | Med. | Prev. | Intensity | Intensity | Intensity | Intensity |
| Intensity | Intensity | Intensity |
| Discocotyle sagittata | 1.7 | 1.0 | 35.3 | 3.4 | 2.0 | 77.8 | 6.2 | 6.0 | 90.9 | 6.1 | 7.0 | 82.4 | 8.5 | 5.0 | 84.6 | 8.8 | 5.5 | 100 |
| Phylodistomum umbilae | 2.2 | 2.0 | 27.8 | 5.2 | 3.5 | 100 | 10.3 | 10.0 | 90.0 | 7.6 | 6.0 | 81.3 | 8.5 | 4.5 | 60.0 | 4.3 | 2.0 | 70.0 |
| Crepidostomum farionis | 2.0 | 2.0 | 25.0 | 3.0 | 3.0 | 16.7 | 3.0 | 3.0 | 50.0 | 2.8 | 3.0 | 38.5 | 3.4 | 3.0 | 100 | 1.7 | 1.0 | 37.5 |
| Crepidostomum metoecus | 10.7 | 5.0 | 100 | 16.2 | 14.5 | 100 | 17.4 | 8.0 | 100 | 21.3 | 20.0 | 100 | 27.8 | 15.0 | 100 | 29.4 | 21.5 | 100 |
| Diplostomum sp. | 17.3 | 12.0 | 88.2 | 59.6 | 49.0 | 100 | 54.8 | 54.5 | 100 | 60.4 | 55.5 | 100 | 66.7 | 55.0 | 100 | 111.7 | 86.0 | 100 |
| Proteocephalus sp. | 1.0 | 1.0 | 5.9 | 0.0 | 0.0 | 0.0 | 3.0 | 3.0 | 18.2 | 3.3 | 3.0 | 17.6 | 1.5 | 1.5 | 15.4 | 11.5 | 11.5 | 16.7 |
| Cyathocephalus truncatus | 1.3 | 1.0 | 23.5 | 1.7 | 2.0 | 33.3 | 3.5 | 3.0 | 33.3 | 4.6 | 2.0 | 47.1 | 1.4 | 1.0 | 38.5 | 6.0 | 3.0 | 58.3 |
| Dibothrioccephalus ditremus | 1.8 | 2.0 | 29.4 | 18.5 | 18.5 | 22.2 | 11.0 | 8.0 | 27.3 | 23.3 | 7.0 | 60.9 | 10.0 | 5.0 | 44.4 | 44.4 | 2.0 | 50.0 |
| Dibothrioccephalus dendriticus | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 1.0 | 1.0 | 5.60 |
| Capillaria sp. | 1.0 | 1.0 | 17.6 | 2.0 | 2.0 | 42.9 | 2.8 | 2.5 | 33.3 | 2.7 | 2.0 | 50.0 | 1.7 | 2.0 | 42.9 | 3.5 | 3.5 | 20.0 |

Figure 3. Stomach content composition (volume %) of brown trout in the length-classes 10–19.9 cm, 20.0–29.9 cm, and 30.0–39.9 cm (almost corresponding to the age-classes 3–4, 5–7, and 8–15 winters, respectively), from the lake Øvre Heimdalsvatn, sampled in June (black), July (grey) and September (patterned) 2011.

**Cyathocephalus truncatus**

The mean and median intensity of infection with *C. truncatus* was low in 2011, with a prevalence varying from 23.5 % in age-class 3–4 to a maximum of 58.3 % in age-class 11–15 winters (Table 3). There was a significant linear relationship between age-classes and the relative density of *C. truncatus* (linear regression, age: $R^2=0.69$, $df=6$, $F=11.15$, $p<0.05$), and correspondingly between length-classes and relative density (linear regression; $R^2=0.71$, $df=5$, $F=9.65$, $p<0.05$) (Figure 4). The positive relationship between relative volume of *G. lacustris* in stomach contents of brown trout age-classes 3 to 9 winters and the relative density of *C. truncatus* was also significant (linear regression; $R^2=0.82$, $df=6$, $F=23.26$, $p<0.01$) (Figure 5).

**Proteocephalus sp**

*Proteocephalus sp.* was primarily found in June, with only one specimen observed in July and none in September. The mean and median intensity of infection was low in all age-classes, as was the prevalence (Table 3).

**Crepidostomum spp**

*C. metoecus* was recorded in all examined trout and had a high mean and median intensity compared to *C. farionis*, which also had a much lower prevalence in all age-classes (Table 3). The positive relationships between age-classes/length-classes, respectively, and relative density of *Crepidostomum spp.* were both significant (linear regression, age: $R^2=0.88$, $df=6$, $F=37.21$, $p<0.01$; length: $R^2=0.66$, $df=5$, $F=7.83$, $p<0.05$) (Figure 6). There was no significant relationship between the relative volume of *G. lacustris* in stomach contents and relative density of *C. metoecus* in the age-classes 3–9 winters ($R^2=0.34$, $df=6$, $F=2.62$, $p>0.05$). Neither was there a significant relationship between relative volume percent of *G. lacustris* + mayflies in stomach contents (both expected intermediate hosts) and relative density of *Crepidostomum spp.* in age-class 3–9 winters ($R^2=0.54$, $df=6$, $F=6.02$, $p<0.05$).

**Diplostomum sp.**

*Diplostomum sp.* had a high mean and median intensity, and with a 100 % prevalence in all age-classes ≥5 winters (Table 3). The positive
classes in 2011, but with highest intensities in age-class 11–15 winters, and with highest prevalence in age-class 7–8 winters (Table 3).

**Discocotyle sagittata**
The only helminth species with a direct life cycle, *D. sagittata*, was common on the trout gills in 2011, with an increase in both prevalence and mean and median intensity with increasing age of the trout (Table 3). The positive linear regressions between age-classes/length classes and relative density, were both significant (age: $R^2=0.89$, df=6, $F=40.78$, $p<0.01$; length: $R^2=0.67$, df=5, $F=8.36$, $p<0.05$) (Figure 6).

**Phyllodistomum umblae**
The kidney fluke *P. umblae* was common in 2011, with a prevalence varying between 27.8 and 100% in the examined age-classes (Table 3). The mean and median intensity was however relatively low (Table 3). The relative density peaked in age-class 6 and 7, and in length-class 20.0–24.9 and 25.0–29.9 cm (Figure 6). There was no significant linear relationship between age-class/length-class and relative density of this species.

**Capillaria sp.**
The mean and median intensity of *Capillaria sp.* was low in all age-classes in 2011, but with highest intensities in age-class 11–15 winters, and with highest prevalence in age-class 7–8 winters (Table 3).

**DISCUSSION**
When compared to the study performed by Lien (1976), the helminth species composition in brown trout in the lake, Øvre Heimdalsvatn, was almost unchanged from 1969–72 to 2011 (Table 4). However, the relative densities (abundance) of some species have changed markedly. The most obvious differences are the substantial increase in *D. ditremus* and *Diplostomum sp.*, i.e., two species which have birds as their final hosts. Some species have had a marked decrease in relative density, especially *Crepidostomum spp.*, but also *C. truncatus* and *Proteocephalus sp.* The other species, *P. umblae*, *D. sagittata*, and *Capillaria sp.* did not show any marked differences in occurrence between 1969–72 and 2011, except the substantial decrease of *P. umblae* in the youngest trout (Table 4).

For parasite species of fish which are transmitted by ingestion of infected invertebrates, there seems to be a close association between infection intensity of a specific species and consumption of the corresponding intermediate host (Knudsen 1995; Knudsen et al. 1996). Differences in parasite community structure and abundance may therefore be closely related to differences in food niches of fish (Knudsen et al. 1997). It may therefore be instructive to relate changes
in parasite abundance in brown trout in the lake Øvre Heimdalsvatn to possible changes in abundance of infected intermediate hosts, in addition to changes in the fish community and the number of piscivorous birds foraging in the lake. Several of these factors may affect the parasite abundance of brown trout in this lake. The brown trout summer diet changed considerably between the period 1969–72 and the years previous to 2011, with a major decline in *L. arcticus*, and with fish becoming an important part of trout diet, mainly European minnow but also brown trout (Lien 1978; Borgstrøm et al. 2010). Both brown trout and European minnow were completely absent from trout diet during the years 1969–72 despite extensive sampling throughout the year (Lien 1978).

The most striking change in infection was the increased occurrence of *D. ditremus*, a species which was almost absent from brown trout in the lake during the period 1969–72, whereas in 2011 it was found in fish within all age-classes (Table 4), and with very high numbers

![Graphs showing parasite density over age and length classes.](image)

**Figure 4.** The relative density of *Crepidostomum* spp., *Diplostomum* sp., *Phyllobothrium* umblae, and *Discocotyle* sagittata in the seven age-classes (3–≥9 winters), and in six length-classes (from 10.0–14.9 cm to 35.0–39.9 cm) of brown trout sampled in the lake, Øvre Heimdalsvatn, in June, July, and September 2011.
in some older fish. This was also documented in a follow-up study in 2012 (Borgstrøm et al. 2017). *D. ditremus* has copepods and salmonids as first and second intermediate hosts, respectively, with fish-eating birds as final hosts (Henricson 1977; Andersen & Gibson 1989). Copepods were more or less absent in brown trout summer diet, both in 1969–72 (Lien 1978) and 2011. In late autumn 1969–72, however, copepods were part of the diet (Lien 1978), and copepods were found in summer diet of small brown trout (8.0–15.9 cm) in 1996 as well (Museth et al. 2010). Seasonal and annual variations in copepod feeding and thereby variations in infection pattern might therefore occur. Nevertheless, during the years 1969–72 when brown trout had copepods as part of their diet, the relative density of *D. ditremus* was zero in age-classes 3–6, and only 0.1 in older brown trout, indicating that the frequency of infected copepods in the diet had been very low compared to the situation in 2011. The very high numbers of *D. ditremus* in some older fish indicate a feeding specialization on copepods, as seen in Arctic char (Knudsen et al. 1996). Alternatively, the high numbers of *D. ditremus* in single fish may be caused by piscivory. If an infected brown trout is eaten by a piscivorous con-specific, the plerocercoids may survive and re-establish in the new intermediate host. Cannibalism seems, however, to be seldom in Øvre Heimdalsvatn, with only 1.4 brown trout juveniles (with length around 10 cm) recorded per 100 examined brown trout stomachs in the period 1993–2005 (Borgstrøm et al. 2010). In Arctic char, large and old individuals are often heavily infected by *D. ditremus* (Hammar 2000; Borgstrøm et al. 2015), most likely due to cannibalism. In the small lake, Nordre Borgdam, Spitsbergen, one individual was infected with around 2000 plerocercoids of *D. ditremus*, and at the same time the large individuals were cannibals (Borgstrøm et al. 2015), thus indicating a possible transfer of larvae from char to char. Similar transmissions may occur from trout to trout as well, although it has been demonstrated that *D. ditremus* has a poor ability to re-establish in fish (Halvorsen & Wissler 1973). In 2011, young brown trout (age-class 3–4) in Øvre Heimdalsvatn had a low relative density of *D. ditremus*, and most examined small fish were uninfected, and accordingly, the very high number of *D. ditremus* plerocercoids in some large brown trout is hard to explain by cannibalism. There is however a third possibility of transfer from copepods to brown trout; European minnow in the lake also feed on copepods (Museth et al. 2010), and the piscivorous trout might obtain *D. ditremus* by consumption of minnows which have infected copepods in their stomachs. In addition, predation on minnows in the lake Øvre Heimdalsvatn is substantial and seems to be a major mortality factor for sexually mature European minnow (Museth et al. 2003).

According to Lien (1981) the number of European minnows >65 mm was 14,500 and 23,000 in 1977 and 1978, respectively. In 1999 and 2000, the estimated number had increased to 89,000 and 35,000, respectively (Museth et al. 2002). During the initial phase of European minnow population build-up, in the years 1969–72, the population number was probably much lower than even in 1977–78. The large population increase in minnows may have made the lake a more favourable foraging locality for fish-eating birds. The frequent observations of the black-throated diver during the last decades before 2011 are in strong contrast to only two observations during the years 1968–72 (Lien & Nydal 1973). Observations of mergansers have also increased, with nesting in the proximity of the lake. The increased numbers of these fish-eating birds may thus have contributed to the present high infection intensities of *D. ditremus* in brown trout, by an increased production of *Dibothriocephalus* eggs ending up in the lake via bird faeces.

Likewise, the increased number of fish-eating birds in Øvre Heimdalsvatn may also have affected the abundance of the eyeflake, *Diplostomum* sp. in brown trout. The relative density was considerably higher in all year-classes compared to 1969–72 (Table 4), with especially high infection intensities in old fish. Fish-eating birds are the final hosts, while the snail species, *Radix* spp., and a variety of fish are first and second intermediate hosts of *Diplostomum* spp., respectively (Palmieri et al. 1976). *R. balthica* is the only suitable first intermediate host in Øvre Heimdalsvatn, and it is mainly located in the shallow part of the lake, although it can be found sporadically at all depths (Brittain 1978). Both brown trout and European minnow may serve as intermediate hosts for *Diplostomum* species (Kennedy 1974), and through the establishment of European minnow in the lake, the abundance of secondary intermediate hosts has increased considerably since 1969–72 (Museth et al. 2002). Buck & Lutterschmidt (2017) studied metacercarial abundance of a trematode parasite (*Posthodiplostomum minimum* Dubois, 1936) in two species of centrarchid fishes (Bluegill *Lepomis macrochirus* Rafinesque, 1819 and redbreast sunfish *L. auritus* Linnaeus, 1758), and they found that host density was negatively associated with parasite abundance, i.e., a demonstration of an encounter-dilution effect. With the presence of minnows in the lake, Øvre Heimdalsvatn, the number of ‘targets’ (intermediate hosts) for *Diplostomum* cercariae will have increased manifold, leading to a possible dilution of the metacercaria encountering each intermediate host, including brown trout, thereby giving a reduced abundance of metacercaria in the brown trout. However, we found the opposite, a marked increase in infection intensity compared with the historical data (Table 4). This increase can be explained both by the increase in number of bird hosts foraging in the lake, and by an increased availability of infected fish. Metacercariae of *Diplostomum* may reduce the vision of fish by inducing catatarch formation (Karvonen et al. 2004), which may alter fish escape responses and cryptis, and also change the shoaling behaviour of infected fish (Seppälä et al. 2008). European minnows typically display such shoaling behaviour. Reduced shoaling of *Diplostomum*-infected fish may predispose them to predation by birds, and thus provide a potential mechanism to enhance the transmission of the parasite to its final hosts, as suggested by Seppälä et al. (2008) in a study of rainbow trout (*Oncorhynchus mykiss* Walbaum, 1792) infected by *D. spathaceum*. In this way, the establishment of a large population of European minnow in the lake Øvre Heimdalsvatn, may have facilitated the transmission of metacercariae from fish to the final hosts, also over-ruling the effect of a possible transmission-dilution, resulting in more egg-producing individuals of *Diplostomum* sp. in the fish-eating birds.

The occurrence of *Diplostomum* sp. in brown trout in Øvre Heimdalsvatn showed a significant positive relationship with age and length of fish, which may be due to exposure to infection over time, as well as a larger body surface of older fish, increasing both contact and penetration possibility for the cercariae of *Diplostomum* sp. Free-living cercariae penetrate the skin or gills of fish (Karvonen et al. 2003), and since older fish may have been exposed for a longer time, the infection may build up. The free-swimming cercariae do not actively search for hosts, but they are often found in concentrations around infected snails (Karvonen et al. 2003), and risk of infection is therefore highest in littoral areas.

In contrast to the increased occurrence of *D. ditremus* and *Diplostomum* sp., the mean intensity of *Crepidostomum* spp. was substantially lower in all age-classes of brown trout relative to 1969–72 (Table 4), indicating a decreased intake of infected intermediate hosts. Fitzgerald & Mulcahy (1983) found the highest occurrence of these species during winter - early spring, and the lowest during
midsummer. Crepidostomum spp. have Radix and Pisidium as first intermediate hosts, mayflies (Ephemeroptera) and amphipods (Gammarus) as second intermediate hosts, and trout as a final host (Awachie 1968; Petkevičiūtė et al. 2018). In a study from North Wales, Awachie (1968) found Radix peregrina (Müller, 1774) as the first intermediate host of C. motococcus, the most numerous of the two Crepidostomum species infecting brown trout in Øvre Heimdalsvatn, and Gammarus pulex (Linnaeus, 1758) was described as the second intermediate host. Petkevičiūtė et al. (2018) referred however to a study where Pisidium was recorded as the first intermediate host, and nymphs of the mayfly genus Ameletus Eaton, 1865 as second intermediate host of C. motococcus. Thus, most probably several species of molluscs, gastropods, and mayflies are candidates for first or second intermediate hosts of the Crepidostomum species. The cercariae emerge from the first intermediate hosts and penetrate mayfly nymphs or amphipods, which in turn must be eaten by trout to complete their life cycle. Since Crepidostomum spp. might survive about one year in its final host (Thomas 1958), the significant relationship between age-class/length-class and relative density of Crepidostomum in brown trout from Øvre Heimdalsvatn in 2011 is probably a reflection of the total consumption of intermediate hosts related to fish size. The density of the trout population in Øvre Heimdalsvatn has changed little (Borgstrøm et al. 2010), as also documented in the present study. The infection pattern of C. truncatus is related to the seasonal consumption of G. lacustris (Amundsen et al. 2003), which also supports our finding in Øvre Heimdalsvatn, with a positive relationship between volume percent of G. lacustris in the stomach content of brown trout and the relative density of C. truncatus. The relative density of this species has nevertheless declined from the period 1969–71 (Table 4), indicating a reduced consumption of G. lacustris, however, the sampling months

Which of the intermediate host species that are involved in the life history of the Crepidostomum species in Øvre Heimdalsvatn is still an open question.

A change in availability of intermediate hosts (bivalves) may also explain the change in infection pattern of P. umblae. This species had its highest relative density in the youngest age-classes during 1969–72 (Table 4), while in 2011 the highest relative density was found in age-class 6 winters (Table 4). Young brown trout in the lake have a habitat mostly confined to the shallow littoral where European minnow also occur (Museth et al. 2002). The large minnow population may have reduced the density of the intermediate hosts in this habitat, but not in deeper parts of the lake occupied by larger brown trout.

Cyathocephalus truncatus is another species which has the amphipod, G. lacustris, as its intermediate host, and fish as the final host (Vik 1958). Although a reduction in density of G. lacustris in the littoral of the lake has been documented (Næstad & Brittain 2010), G. lacustris has still remained common in brown trout summer diet (Borgstrøm et al. 2010), as also documented in the present study. The infection pattern of C. truncatus is related to the seasonal consumption of G. lacustris (Amundsen et al. 2003), which also supports our finding in Øvre Heimdalsvatn, with a positive relationship between volume percent of G. lacustris in the stomach content of brown trout and the relative density of C. truncatus. The relative density of this species has nevertheless declined from the period 1969–71 (Table 4), indicating a reduced consumption of G. lacustris, however, the sampling months

Table 4. Summary of the helminth fauna of brown trout from the lake Øvre Heimdalsvatn, as relative density (abundance) of the different helminth species recorded in the age-classes 3–4, 5, 6, and ≥7 winters in 1969–72 (Lien 1976), and in the present study from 2011. Numbers in parentheses represent total number of examined fish.

| Helminth species | 1969 – 1972 | 2011 |
|------------------|------------|------|
|                  | Age-class (winters) | Age-class (winters) |
|                  | 3–4 (69) | 5 (84) | 6 (78) | ≥7 (98) | 3–4 (22) | 5 (16) | 6 (15) | ≥7 (59) |
| Discocotyle sagittata | 0.5 | 1.7 | 2.4 | 4.7 | 0.6 | 2.6 | 5.6 | 6.8 |
| Phylysodistomum umbilae | 13.7 | 10.8 | 9.2 | 6.3 | 0.6 | 5.2 | 9.3 | 5.0 |
| Crepidostomum spp. (two species) | 147.0 | 229.0 | 234.0 | 233.0 | 11.2 | 16.7 | 18.9 | 26.5 |
| Diplostomum sp. | 12.0 | 10.8 | 5.9 | 17.2 | 17.0 | 59.6 | 54.8 | 78.4 |
| Proteocephalus sp. | 13.5 | 11.8 | 11.4 | 7.8 | 0.1 | 0.0 | 0.5 | 0.9 |
| Cyathocephalus truncatus | 3.1 | 4.9 | 7.3 | 9.7 | 0.3 | 0.6 | 1.2 | 2.0 |
| Dibothriocephalus ditremus | 0.0 | 0.0 | 0.0 | 0.1 | 0.5 | 4.1 | 3.0 | 13.7 |
| Dibothriocephalus dentriticus | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.02 |
| Capillaria sp. | 1.1 | 1.6 | 0.8 | 0.6 | 0.2 | 0.9 | 0.9 | 1.0 |
| Euastronyxides sp. | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
have to be considered, especially because *C. truncatus* spend only a few months in its final host (Vik 1958). The seasonal infection pattern of *C. truncatus* in 1970–71, with low abundance during the summer months June - August (Figure 7) may indicate a higher loss of the parasite during these months, or a lower consumption of *G. lacustris*, and the results from 2011 are consequently not directly comparable with the 1969–72 data.

The monogenean, *Discocotyle sagittata*, the only recorded species with a direct life history without intermediate hosts, has nearly the same relative density in 2011 as in 1969–72 (Table 4). In contrast to *C. truncatus*, this monogenean has a temperature dependent transmission (Rubio-Godoy & Tinsley 2008), and the species is therefore able to reproduce and transmit most effectively during the warmer months. During the period 1969–72, samples were taken throughout the year (Lien 1976), whereas the present study was confined to the summer months. The somewhat higher relative density of *D. sagittata* in 2011 may therefore be due to the sampling procedure for this species as well. However, the density of old (and large) brown trout has been considerably higher in the years previous to 2011 compared to the years previous to 1969–72 (Jensen 1977; Borgstrøm et al. 2010). The positive relationship between age and length of brown trout and number of *D. sagittata*, together with a higher abundance of large fish, may therefore result in an increased production of larvae, and thereby also an increased infection pressure.

The apparent decrease in the abundance of *Proteocephalus* sp. (Table 4), may likewise be related to the seasonal occurrence, with low infection intensity in brown trout during the summer months, as documented by Lien & Borgstrøm (1973), suggesting that the low numbers of *Proteocephalus* sp. (similar to *C. truncatus*) present during the summer of 2011 was mainly due to the loss of worms in spring - early summer.

Temperature may greatly influence development and transmission of parasite species, such as *Diplostomum* sp. (Hakalahiti et al. 2006). An increase in annual mean temperatures in Norway from the 1960s to 2010 (Hanssen-Bauer et al. 2017) may indicate a general temperature increase during the summer months in the lake Øvre Heimdalsvatn. However, the long-term development in mean June - August air temperatures, as measured at the nearest meteorological station (Skåbu), shows no increasing trend from the period 1969–72 to the period 2006–09 (https://seklima.met.no). With a high correlation between air temperatures at Skåbu meteorological station and summer temperatures in the outlet river of Øvre Heimdalsvatn (Kvambekk and Melvold 2010), the observed changes in infection intensity and prevalence of *Diplostomum* sp., as well as other helminth species of brown trout in the lake, can hardly be related to changes in summer temperatures.

**CONCLUSIONS**

The long-term changes of the helminth fauna of brown trout in the lake Øvre Heimdalsvatn may be explained by factors influencing the life cycle of the different species, especially: (i) brown trout diet and the availability of invertebrate prey functioning as intermediate hosts, and (ii) density of fish-eating birds as final hosts for parasites with fish as an intermediate host. The most striking difference has been the substantial increase in infection intensities of the two species which have brown trout as an intermediate host and fish-eating birds as final hosts (*Dibothriocephalus latreillei* and *Diplostomum* sp.), and a marked decrease in others, especially the *Crepidostomum* species which have several invertebrates, including *G. lacustris*, mayflies, and molluscs, as intermediate hosts. These invertebrates also serve as important food items for both brown trout and European minnow, and with establishment of a large minnow population, the total predation pressure on invertebrates in the littoral area may have increased, resulting in reduced availability of intermediate hosts. Some of the parasite species show small changes in infection intensities, but due to known seasonal variations in occurrence, and sampling in 2011 restricted to the months June, July, and September, conclusions regarding long-term changes are not possible to draw for these species. The only helminth with a direct life cycle was the monogenean, *Discocotyle sagittata*. Since this species showed a positive relationship with trout size, the small increase in relative density may be related to a greater abundance of older and larger fish in 2011 compared to 1969–72, and thereby an increased reproduction success of this monogenean. In conclusion, changes in the lake ecosystem of Øvre Heimdalsvatn after establishment of the European minnow are likely to have played a decisive role in the observed changes in the parasite fauna of the brown trout.

**ACKNOWLEDGEMENTS**

Special thanks are given to Professor Philip D. Harris formerly at the Natural History Museum, University of Oslo, for help with the parasite identification, as well as his participation during part of the field work. We also acknowledge Professor Jan T. Lifjeld at the Natural History Museum, University of Oslo, for information regarding the bird fauna at Øvre Heimdalsvatn, and the Natural History Museum, University of Oslo, for permission to use the University’s field station. We greatly acknowledge three anonymous referees for their constructive criticism and many suggestions for improving earlier versions of the manuscript.

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Editorial responsibility: Jan Grimsrud Davidsen.

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