Studies on subalpine forests of Hamta Pass area in Himachal Pradesh, India with a focus on Betula utilis populations

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The present study was conducted in Hamta Pass area of Kullu district, Himachal Pradesh, India with a focus on Betula utilis populations. Totally 16 populations of B. utilis representing four habitats and three aspects were studied. The maximum sites were represented by moist and moist, shaded habitats with northwestern aspect. Totally 188 plant species representing trees, shrubs and herbs were recorded. Acer acuminatum, Abies pindrow, Prunus cornuta and Quercus semecarpifolia were the major associated species of B. utilis in the subalpine zone of Hamta Pass. Based on importance value index five tree communities, namely A. acuminatum, A. pindrow, B. utilis, Q. semecarpifolia and B. utilis–P. cornuta mixed were identified. Among the communities, total density of trees, shrubs and herbs was recorded from 160 to 270, 300 to 515 and 21 to 33 individuals m⁻² respectively. The total basal area recorded was 6.94–42.10 m² ha⁻¹, species richness 15–127, and species diversity for trees 0.4–0.9, shrubs 0.0–2.1 and herbs 1.7–4.2. The density and regeneration (i.e. seedlings and saplings) of B. utilis in most of the populations revealed that this species will continue to grow in the area. However, continued anthropogenic activities, climate change and other factors may cause population depletion in the study area.

Keywords: Betula utilis, regeneration, species diversity, species richness, subalpine forests, total basal area.

Introduction

The unique topography, climatic conditions, diverse habitats and large altitudinal range of the Indian Himalaya Region (IHR) constitute an important part in global biodiversity hotspots¹. The mountain ecosystem harbours diverse biological communities and a high level of endemism². The Northwestern Himalaya is irregular and disturbed by valleys and plateaus with great floristic diversity due to altitudinal variations³. There is exorbitant biodiversity in the Himalayan ecosystems, which are recognized for their provisioning, cultural, regulating and supporting services to both upland and lowland inhabitants⁴. However, due to various anthropogenic and changing climate scenarios, these ecosystems and the services provided by them are being severely affected, causing loss of biodiversity and reduction in ecosystem services. Subalpine forests share the elements of high alpine and low temperate zones in the Himalaya, as they form a transition zone between temperate forest and alpine meadows⁵. The biodiversity components (i.e. vegetation structure, composition and function) of the subalpine and alpine ecosystems are severely affected by heavy grazing, overexploitation of woody species for fuel by herdsmen and non-timber forest products (NTFPs), including medicinal plants⁶.

The rapid geo-climatic variations are the basis of different environmental gradients and result in vegetation and diverse plant community types⁷. Habitat loss, fragmentation, overexploitation, alien species invasion and global climate change are the threats to biodiversity and the ecosystem⁸. These threats are also responsible for the decreasing trend of native species, upward shifting of species and the opportunity for proliferation of invasive and exotic species with low economic value⁹. Other than anthropogenic and climatic impacts, topographical factors (i.e. altitude, habitat and aspect) control the distribution patterns of vegetation in the Himalayan ecosystems¹⁰. In the current scenario, subalpine forests are the subject of interest for many researchers and are being studied for the impact of global climate change worldwide.

Betula utilis D. Don (bhojpatra or Himalayan silver birch) is a broadleaved, deciduous, ecologically and economically important tree. It represents one of the dominant species of the Himalayan treeline. The high freezing tolerance features enables this species to form a treeline in the Himalaya¹¹. This species is found in major association with Abies pindrow, Abies spectabilis, Quercus semecarpifolia, Prunus cornuta, Acer acuminatum, Sorbus foliolosa, Pinus wallichiana and Rhododendron campa-nulatum. The bark is the characteristic feature of B. utilis.
due to its shining, reddish-white appearance having both ethnobotanical\cite{12,13} and phytochemical properties like anti-cancerous, anti-HIV, antioxidant, antimicrobial and anti-fertility activities\cite{14-16}.

In general, systematic studies, i.e. qualitative and quantitative assessment of subalpine forests, including *B. utilis* have been carried out by various workers\cite{17-33} in the IHR. However, focused studies on *B. utilis* populations have not been carried out so far. Therefore, in view of the various anthropogenic pressures and changing climate scenario, assessment of floristic diversity of the subalpine forests, including *B. utilis* populations becomes utmost important.

**Materials and methods**

**Study area**

The present study was conducted in Hamta Pass area of Northwestern Himalaya of Kullu district, Himachal Pradesh, India (Figure 1). The subalpine zone in the present study area is dominated by broadleaved deciduous species (i.e. *Q. semecarpifolia*, *P. cornuta*, *A. Acuminatum*, *S. foliolosa*, *Corylus jacquemontii* and *B. utilis*) and evergreen coniferous species (i.e. *A. pindrow*, *A. spectabilis* and *P. wallichiana*). The dominant shrub species are *Rhododendron campanulatum*, *Rhododendron anthropogon*, *Salix denticulata*, *Rosa macrophylla*, *Rosa sericea*, *Smilax vaginata*, *Spiraea bella*, *Juniperus indica* and *Lonicera spp.*

**Vegetation assessment in B. utilis forests**

Extensive and intensive field surveys were conducted during July and August 2018. For the sampling of floristic diversity of *B. utilis* forests, each accessible habitat and aspect was selected and assessed. Based on physical characters and dominance of the vegetation, habitats were identified. Site characteristics, i.e. latitude, longitude, altitude, slope and aspect of each site representing *B. utilis* populations were recorded with the help of global positioning system (GPS), Abney level and compass. For quantitative assessment, a plot of 50 m × 50 m was laid randomly. Within this plot, trees, saplings and seedlings were sampled by laying ten quadrats of 10 m × 10 m; shrubs, 20 quadrats of 5 m × 5 m; and herbs, 20 quadrats of 1 m × 1 m (refs 33–35). Circumference at breast height (cbh), i.e. 1.37 m and diameter at 10 cm from the base for each tree and seedling were recorded for considering individuals as trees (cbh ≥ 31.5 cm), saplings (cbh 10.5–31.4 cm) and seedlings (cbh < 10.5 cm). Shrubs were considered as woody species having several branches arising from their base, and herbs were considered as those species having aerial parts surviving for one season, though their underground parts, i.e. roots/rhizomes/bulbs, etc. may remain alive during other seasons. From each site, samples of each species were collected and identified with the help of local and regional flora\cite{17-21}.

The data collected during field surveys were analysed in MS-Excel following standard ecological methods\cite{36-38}. 

\begin{figure}
\centering
\includegraphics[width=\textwidth]{Figure1.png}
\caption{Map showing the sites of *Betula utilis* populations in Himachal Pradesh, India.}
\end{figure}
On the basis of importance value index (IVI), different communities were identified. Thereafter, data from different populations were pooled to calculate community averages in terms of density, total basal area (TBA) and Shannon–Wiener information index of species diversity \((H')\)\(^{39}\). The total count of species was considered as species richness. The diameter size class was employed to measure the population structure of tree species. Ten random size classes (<10.5 cm for seedlings, 10.5–31.4 cm for saplings, 31.5–41.5, 41.6–51.5, 51.6–61.5, 61.6–71.5, 71.6–81.5, 91.6–100.5 and >100 cm) of individuals tree species were formed. The total number of individuals was calculated from each quadrat for each tree species and relative density was calculated as a percentage of total number of individuals in all size classes.

**Assessment of physico-chemical properties in B. utilis forests**

Soil samples were collected from each studied site within each plot of 50×50 m (15–20 cm depth). Five soil samples, four from the corners and one from the centre of each plot were collected, pooled and mixed properly to make a composite sample. The air-dried soil samples were assessed for further tests and analysis. Soil pH was measured using a pH meter in 1:5 mixture of soil and distilled water, moisture content was recorded as % difference in fresh and dry soil weight, % organic carbon and organic matter were analysed as described by Walkley and Black method\(^{30}\), available nitrogen by kjeldahl method\(^{31}\), available phosphorus by Olsen’s extraction method\(^{42}\) and available potassium by flame photometer\(^{33}\).

**Numerical and statistical analysis**

All numerical analysis was done in MS-Excel. Regression analysis was performed to estimate the relationship among altitude and ecological parameters. Pearson correlation coefficient was calculated between various ecological parameters and soil parameters using STATISTICA-8.

**Results and discussion**

### Site and habitat characteristics

| Site | Altitude (m) | Latitude (N) | Longitude (E) | Habitat | Slope (°) | Aspect | Community |
|------|--------------|--------------|---------------|---------|----------|--------|-----------|
| 1    | 3057         | 32°12'52.32" | 77°14'18.42" | Moist   | 50       | NW     | BU        |
| 2    | 3109         | 32°13'42.54" | 77°14'08.82" | Bouldery| 65       | NW     | AP        |
| 3    | 3150         | 32°14'56.94" | 77°14'41.16" | Shady moist | 55 | SW     | AP        |
| 4    | 3155         | 32°15'03.60" | 77°14'44.22" | Moist   | 55       | NW     | BU        |
| 5    | 3171         | 32°12'55.14" | 77°14'30.72" | Moist   | 60       | NW     | BU        |
| 6    | 3190         | 32°14'54.46" | 77°15'05.94" | Shady moist | 40 | W      | BU        |
| 7    | 3230         | 32°14'58.92" | 77°14'54.84" | Moist   | 60       | W      | AA        |
| 8    | 3258         | 32°15'52.02" | 77°15'13.50" | Moist   | 40       | NW     | BU        |
| 9    | 3310         | 32°13'45.00" | 77°14'20.94" | Shady moist | 50 | W      | QS        |
| 10   | 3317         | 32°13'02.58" | 77°14'54.78" | Moist   | 35       | NW     | BU        |
| 11   | 3364         | 32°13'18.60" | 77°14'55.74" | Shady moist | 40 | SW     | BU        |
| 12   | 3411         | 32°15'46.26" | 77°15'25.08" | Rocky   | 50       | W      | BU        |
| 13   | 3505         | 32°13'45.78" | 77°14'35.34" | Bouldery | 60 | W      | BU        |
| 14   | 3573         | 32°16'01.50" | 77°15'06.54" | Rocky   | 40       | NW     | BU        |
| 15   | 3760         | 32°13'57.06" | 77°15'01.92" | Rocky   | 30       | SW     | BU-PC     |

BU, Betula utilis; AP, Abies pindrow; QS, Quercus semecarpifolia; AA, Acer acuminatum; PC, Prunus cornuta.

The studied sites were assessed from four habitats, i.e. moist, shady moist, rocky and bouldery, and three aspects, i.e. north-west, west and southwest between 32°12'52.32"–32°15'57.06" N lat. and 77°14'18.42"–77°15'01.92" E long. and 3047–3760 m amsl altitude. Table 1 shows the habitat characteristics. The dominant shady, moist and moist habitats were observed in mixed B. utilis forest, while the pure B. utilis stands were found in rocky and bouldery habitats. They formed large stashes of monodominant forests in the NW aspect, and a discontinuous and abrupt treeline in the present study sites. In Trans-Himalaya treeline, B. utilis is dominant in the north-facing slope\(^{44}\). The soil receives moisture from seasonal snow on the hilltop and winter snow is the main source of moisture. The western and south-western facing slopes are quite dry and B. utilis populations exist in scattered pockets. Soil moisture and aspect are the governing and controlling factors for B. utilis distribution in the sub-alpine zone of Manag valley, Nepal Himalaya. The north-facing slopes receive a small part of monsoon rainfall, therefore, snowmelt-water from the high mountains is the main source of soil moisture\(^{44}\).

### Floristic composition in B. utilis forests

The field study was conducted in the subalpine zone of Hamta Pass, which is unexplored in terms of floristic
Table 2. Associated life-forms of *B. utilis* forests in subalpine zone of Hamta Pass area

| Tree Species | Shrub Species | Herb Species |
|--------------|---------------|--------------|
| *Abies pindrow* | *Clematis montata* | *Ainsliaea aperta* |
| *Acer acuminatum* | *Juniperus communis* | *Adiantum venustum* |
| *Betula utilis* | *Juniperus macropoda* | *Bistorta affine* |
| *Prunus cornuta* | *Lonicerat parviflora* | *Anaphalis contorta* |
| *Quercus semecarpifolia* | *Rosa macrophylla* | *Angelica archangelica* |
| | *Rosa moschata* | *Angelica glauca* |
| | *Rosa sericia* | *Aquilegia pubiflora* |
| | *Salix denticulata* | *Caltha palustris* |
| | *Salix lindleyana* | *Dactylorhiza hatagirea* |
| | *Smilax vaginata* | *Delphinium brunonianum* |
| | | *Inula grandiflora* |
| | *Polygonatum verticillatum* | |
| | *Polygonatum cirrhifolium* | |
| | *Sinopodophyllum hexandrum* | |
| | *Smilacina oleracea* | |
| | *Trillium govanianum* | |

Table 3. Community-wise species richness, density, total basal area (TBA) and species diversity index of trees, shrubs and herbs in Hamta Pass area

| Community | Site representation | Species richness | TBA (M² ha⁻¹) | Density | Species diversity index (H') |
|-----------|---------------------|-----------------|--------------|---------|-----------------------------|
| AA        | 1                   | 36 (3T, 2S, 1H) | 17.22        | 260     | 0.9                         |
| AP        | 2                   | 55 (4T, 3S, 48H) | 42.10        | 250     | 0.8                         |
| BU        | 10                  | 127 (5T, 11S, 111H) | 17.36       | 226     | 0.9                         |
| QS        | 2                   | 45 (2T, 2S, 41H) | 6.94         | 160     | 0.7                         |
| BU–PC     | 1                   | 15 (3T, 1S, 11H) | 20.92        | 270     | 0.4                         |

T, Trees; S, Shrubs; H, Herbs.

The tree species richness (5) in the present study is similar to the treeline ecotone of Eastern45 and Central Nepal Himalaya46. High species richness in pure *B. utilis* community might be due to decrease in canopy cover. Similar trends of species richness in pure *B. utilis* community have also been reported from Kedarnath Wildlife Sanctuary (KWS)37. The recorded species richness was higher than the earlier reported value from the Kumaun Himalaya24,48, Hirb and Shoja catchments of Himachal Pradesh33 and subalpine forests of KWS31. The highest species richness in the *B. utilis* community could be due to overlapping of species towards the centre from the highest and lowest elevations49. High forest density and forest cover in the subalpine zone, are a consequence of low light intensity under the forest canopy50. The scanty ground vegetation in subalpine forests is linked with decrease in temperature in this zone. The highest total tree density (270 individuals ha⁻¹), shrub density (515 individuals ha⁻¹) and herb density (33 individuals m⁻²) was contributed by *B. utilis–P. cornuta* mixed community. A total of five tree communities namely, *A. acuminatum*, *A. pindrow*, *B. utilis*, *Q. semecarpifolia* and *B. utilis–P. cornuta* mixed were recorded on the basis of IVI. Maximum ten sites were represented in *B. utilis* community, followed by two sites in *A. pindrow* and *Q. semecarpifolia* each and one site each in *A. acuminatum* and *B. utilis–P. cornuta* mixed communities (Table 3).
Figure 2. Regeneration status of trees in the identified communities based on the density of trees, saplings and seedlings of the species. a, Quercus semecarpifolia community. b, Acer acuminatum community. c, Betula utilis community. d, Betula utilis–Prunus cornuta mixed community. e, Abies pindrow community.

Figure 3. Altitude-wise density of B. utilis trees, saplings and seedlings.

(20.92 m² ha⁻¹) mixed, B. utilis (17.36 m² ha⁻¹), A. acuminatum (17.22 m² ha⁻¹) and pure Q. semecarpifolia (6.94 m² ha⁻¹) communities respectively (Table 3). TBA with increasing altitude showed discontinuous pattern and was maximum at higher elevation. Along with increasing altitude, the growth of trees was stunted and small with lower girth class. TBA and density for trees and shrubs were comparable to earlier studies from temperate and subalpine forests in IHR27,28,31,32,48,51,52.

Species diversity

Community-wise highest species diversity index (H′) for trees was recorded in B. utilis and A. acuminatum communities (0.9 each), followed by A. pindrow (0.8), Q. semecarpifolia mixed (0.7) and B. utilis–P. cornuta mixed (0.4) communities. The highest H′ for shrubs (2.1) was recorded for B. utilis community and lowest in B. utilis–P. cornuta mixed (nil) community. The highest H′ diversity for herbs was recorded in B. utilis (4.2) community, followed by A. pindrow (3.5), A. acuminatum and Q. semecarpifolia (3.3) pure communities and B. utilis–P. cornuta mixed (1.7) communities (Table 3). The H′ values are comparable with those of the earlier studies from temperate and subalpine forests24,26,32,33,48,51,52.

Regeneration pattern of trees in forest communities

Figure 2 a–e shows the regeneration pattern of trees in the identified communities. The regeneration of B. utilis varied with community. Regeneration pattern was observed on the basis of seedlings and saplings density in the forest communities. The regeneration of dominant species in B. utilis community was highest, particularly in sapling layer. Regeneration of dominant species in Q. semecarpifolia, B. utilis–P. cornuta mixed and A. pindrow communities was highest only in terms of seedling layer. While poor regeneration of dominant species was recorded in A. acuminatum and B. utilis pure communities. The poor recruitment of seedlings and establishment of saplings of dominant species in the recorded communities may be due to high tree canopy coverage and lack of microhabitat, high litter accumulation and low light availability affecting the seed germination of B. utilis53.
Altitude-wise regeneration of *B. utilis*

Discontinuous pattern of *B. utilis* trees, saplings and seedlings density was observed along with increasing altitude (Figure 3). The maximum studied *B. utilis* population was recorded with zero sapling density. The gap in the distribution of *B. utilis* seedlings and saplings might be due to high heterogeneity. Other factors responsible for the discontinuous recruitment of seedlings and saplings may be high canopy cover and aspect. Absence of seedlings density in the southwest and west aspects also supports the fact that aspect plays a crucial role in species distribution. In Hamta Pass, the habitat of *B. utilis* in the southwest and west slopes was shady moist due to high canopy coverage.

Regression analysis on density patterns of total trees, *B. utilis*, saplings and seedlings did not show uniform pattern with increasing altitude. The total tree density and *B. utilis* density decreased with increasing altitude due to monodominant forest of *B. utilis* and scanty distribution of the species in the subalpine zone of Hamta Pass (Figure 4). The density of total trees, *B. utilis* and seedlings along the altitude was statistically non-significant at \( P \leq 0.05 \) (Figure 5a–c). However, increase in saplings density along the altitude was statistically significant at \( P \leq 0.05 \) (Figure 5d). Along the increasing elevation, shrubs and herbs density pattern was also statistically non-significant at \( P \leq 0.05 \). The maximum density for both shrubs and herbs was found between 3100 and 3400 m amsl (Figure 5e and f). This may be due to longer distance from pasture land and less disturbance of livestock grazing pressure. In general, herb density increases with increase in elevation; but in the upper subalpine zone of Hamta Pass, assessment was done in August when the area was highly exploited by overgrazing, especially the herb layer that might be the reason for decreasing density with increasing altitude. Also, altitude plays an important role in species distribution and allows specific species to grow.

It was assumed that soil moisture availability affects the regeneration of species in subalpine forests. However, inadequate soil moisture might be responsible for seed germination and recruitment of seedlings and saplings in the Hamta Pass areas, as there are no glaciers. In the treeline ecotone, glaciated snow retreat is the main source of soil moisture content. Seedlings are more sensitive to moisture deficiency than deep-rooted mature plants, which might be attributed to slow initial growth. Seeds are extremely light-weight (0.209 ± 0.019 mg) and are dispersed by wind. Bryophyte mats (mosses) are an excellent substrate for seed germination and establishment due to their high moisture storage capacity; bryophyte mats act as a seed-trapper. However, litter is unfavourable for germination and establishment of *B. utilis* seeds. Thus, lack of suitable microhabitat for seed germination and adverse climatic conditions (i.e. avalanche, snowdrift, low temperature, etc.) and anthropogenic pressures (i.e. overgrazing, camping and hiking) might be responsible for lower recruitment and establishment of seedlings in the subalpine zone of Hamta Pass. However, upward shift in *B. utilis* was not observed in Hamta Pass.

**Size class distribution of tree species in *B. utilis* forests**

The population structures of all tree species, except *A. acuminatum* exhibited similar trends (Figure 4). High accumulation of seedlings and saplings of *Q. semecarpifolia*, *B. utilis* and *A. pindrow* is the characteristic of subalpine forests in Hamta Pass. *A. acuminatum* showed higher accumulation towards tree size class with absence of seedlings and saplings. While *P. cornuta* showed higher accumulation in tree size class more than 100 cm. The sufficient seedlings, saplings and young trees in any population indicate good regeneration status. The environmental conditions in the subalpine zone do not permit many species to regenerate. However, treeline ecosystem is more sensitive to climate change and is responsible for the variations in distribution, growth pattern and seed-based regeneration of species across the treeline. In particular reference to the regeneration of *B. utilis*, the growth of seedlings, saplings and young trees is ambient and resembles reverse J-shaped trends, showing good regeneration in the present study sites (Figure 4).

**Soil properties in *B. utilis* forest**

The edaphic factors, moisture content and nutrients are required for the survival of most of the species. Figure 6 presents altitude-wise soil results. The soil pH was slightly acidic in nature in birch forests across all elevations. Spatial variations in soil properties in forests are
mainly due to the rooting pattern and litter accumulation of the perennial vegetation. The acidic pH in *B. utilis* forest was also reported in subalpine region of Nepal Himalaya. High litter decomposition and organic matter decrease pH in forests. However, pH range from 5.5 to 6.5 is satisfactory for the growth of most of the plant species. Electrical conductivity (EC) and available potassium showed an increase with increasing altitude, while organic carbon and organic matter significantly decreased (Table 4). No significant correlation was observed in moisture content with elevation. But, slight increase in moisture content with elevation might be due to decrease in temperature and increase in tree canopy cover (Figure 6). Soil moisture and temperature are limiting and controlling factors for the significant variation in topographical characters and vegetation.
stands structure and physiognomy of species in treeline. Available P ranged from 0.07 to 0.71 ppm, available K from 31 to 82.37 ppm and available N from 93.33 to 189.40 ppm (Figure 6). Positive significant correlation was observed between moisture content and EC. Soil EC depends on the availability and mobility of ions. The concentration of ions in the soil increases with increase in moisture content and thereby a significant increase in EC. Negative significant correlation was observed between EC and organic carbon and organic matter. The organic matter also showed negative correlation with available soil nitrogen and potassium (Table 4).

Table 5 shows community-wise soil parameters. The highest moisture content (46.1%) and EC (236 μS) were recorded in B. utilis–P. cornuta mixed and A. acuminate pure communities. Organic carbon (3.1%) and organic matter (5.3%) were maximum in A. acuminate, followed by A. pindrow, Q. semecarpifolia, B. utilis and B. utilis–P. cornuta mixed communities. Whereas available phosphorus and potassium were maximum in B. utilis–P. cornuta mixed forest followed by B. utilis pure forest with least recorded in A. acuminate and A. pindrow communities. The highest available nitrogen (150.3 ppm) was recorded in A. acuminate community. Community type may contribute in determining the nutrients cycling in forest ecosystems because of canopy complexity in the mixed forests. The canopy complexity might be responsible for nutrient production and availability. It also produces heterogeneity of soil nutrients in the mixed forest. According to Vesterdal et al., tree species can also be considered as an indicator of soil carbon and nitrogen. Deciduous species have larger forest floor than other forest species for carbon accumulation. Shedayi et al. reported that soil total carbon (STC) and soil total nitrogen (STN) were highest in B. utilis forest, and highlighted the significant impact of vegetation type and litter on STC and STN, while there was no impact of altitude and herbaceous biomass on carbon and nitrogen concentration in the soil.
The anthropogenic activities such as overgrazing, tourism, trampling and camping cause loss of biodiversity. The grazing pressure by livestock (i.e. horse, buffalo and sheep) on subalpine and alpine mats was high (Figure 7). Overexploitation of trees and shrubs for fuel and fodder by the migrants during grazing season affect growth and development of tree species in the subalpine zone. Economically important species such as Trillium govanianum, Angelica glauca and Dactylorhiza hatagirea are extracted by the herdsmen and local inhabitants, which also causes a decrease in the population of species.

**Conclusion and recommendations**

The present study shows floristic richness represented by a total of 188 species of trees, shrubs and herbs distributed over four habitats and three aspects. The populations of the *B. utilis* were dominated by *B. utilis*, *A. acuminatum*, *A. pindrow*, *Prunus cornuta* and *Q. semecarpifolia*. The spatial heterogeneity pattern of saplings and seedlings of dominant tree species was recorded in different community structures. With discontinuous and abrupt treeline; *B. utilis* was the dominant treeline-forming species in Hamta Pass. Overexploitation, deforestation, overgrazing and natural calamities such as erosion, snow drift, forest fires and landslides might be important factors responsible for the recruitment and establishment of seedlings and saplings in the study area. The density and regeneration (i.e. seedlings and saplings) of *B. utilis* in most of the populations revealed that this species will continue to grow in the study area in future as used. However, the continued anthropogenic activities, climate change and other natural factors may cause population depletion in the area. Frequent monitoring of the populations representing *B. utilis* would help in understanding the dynamics of the vegetation and impact of anthropogenic activities, impact of climatic and non-climatic factors on population, and identification of microhabitats for seed germination which may strengthen the conservation and sustainability of *B. utilis* in the near future. Thus, the present study would form the baseline to suggest strategies for the conservation of *B. utilis* populations in natural habitats.

**Figure 7.** Livestock grazing pressure in Hamta Pass area.

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