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The impact of a catastrophic storm event on benthic macroinvertebrate communities in upland headwater streams and potential implications for ecological diversity and assessment of ecological status

Hugh B. FEELEY,1* Stephen DAVIS,1 Michael BRUEN,2 Sean BLACKLOCKE,2 Mary KELLY-QUINN1

1Freshwater Biodiversity, Ecology and Fisheries Research Group (freBEF), UCD School of Biology and Environmental Science, Science Education and Research Centre (West), University College Dublin, Belfield, Dublin 4, Ireland; 2UCD School of Civil, Structural and Environmental Engineering, University College Dublin, Newstead, Belfield, Dublin 4, Ireland

*Corresponding author: hughfeeley@gmail.com

ABSTRACT

Upland headwater streams are dynamic systems, responding rapidly to changes in climatic conditions. This study examined the effects of a catastrophic rainfall event, that occurred on 24 October 2011 on the east coast of Ireland, on the macroinvertebrate community composition and structure of four headwater streams in the river Liffey catchment located in the Wicklow Mountains. The ecological status before and after the storm were also evaluated. The water level and pH of each stream were recorded using continuous monitoring equipment, while rainfall data for the study period were sourced from a local weather station. Benthic macroinvertebrates were investigated before and after the storm event using Surber sampling. Results showed rapid and large increases in water level and significant declines in stream pH in response to intensive rainfall during the storm. The high water levels also caused major physical damage and abrasion in all four streams, that significantly altered instream habitats. The storm event induced significant losses to the richness and/or density of most taxonomic groups, with the exception of the Plecoptera. Furthermore, the overall community composition and structure changed significantly, most likely as a result of physical disturbance, given the relative persistence of acid-sensitive taxa and the relatively short period of harsh acidic conditions (<5 pH). Interestingly however, the ecological status of each of the four study sites, tested using the Small Stream Risk Score (SSRS), the Biological Monitoring Working Party (BMWP) and the Average Score Per Taxon (ASPT) indices, was unaltered by the loss in richness and densities. This was likely a result of the maintenance of plecopteran richness and the absence of organic pollution, thus highlighting the need to develop appropriate indices to assess the ecological status of streams and rivers affected by physical disturbance caused by large storm events. Ultimately, catastrophic storm events in upland headwater streams have potentially major implications for the maintenance of regional macroinvertebrate diversity within affected regions.

Key words: acidification, bioassessment, biodiversity, climate, disturbance, Ireland.

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INTRODUCTION

Upland headwater streams are typically dynamic and changeable habitats, highly susceptible to rapid increases in flow during periods of heavy rainfall (Resh et al., 1988). The frequency, amount and intensity of rainfall, together with antecedent climatic conditions have been shown to influence the magnitude, duration and nature of episodic stream acidification occurring during such high-flow conditions (Ormerod et al. 1989; Kelly-Quinn et al. 1996; Kowalik et al. 2007). This in turn has been shown to affect benthic macroinvertebrate survival in such streams (Tierney et al., 1998; Lepori et al., 2003; Kowalik and Ormerod 2006; Tixier et al., 2009; Feeley et al., 2011). Furthermore, the increased discharges and velocities, and the resultant disturbance and movement of the streambed substrates during high-flow events alter macroinvertebrate community composition and structure within streams, especially during the peaks of high-flow periods (Robinson and Minshall, 1986; Brittain and Eikland, 1988; Resh et al., 1988; Death, 2003). As a result, the interaction between the magnitude, duration and frequency of storm events is a major factor in species survival (Giller et al., 1991; Weatherley and Ormerod, 1991; McCabe and Gotelli, 2000).

The composition and structure of macroinvertebrate communities in dynamic streams is relatively insensitive to natural disturbance cycles (i.e. high flow events) (Resh et al., 1988; Townsend, 1989; Giller et al., 1991). For example, streams affected by episodic acidity (i.e. intermittent periods of low pH) will be dominated by acid-tolerant communities, with some acid-sensitive taxa surviving low pH by burrowing into the streambed and/or drifting downstream, the latter leading to periods of ecological impact (Ormerod et al., 1987; Tixier and Guérol, 2005; Gibbons et al., 2010; Feeley et al., 2011). Nevertheless, the effects of inordinately large storm events are more difficult to predict due to the severity, magnitude, spatial extent and rarity of such disturbances relative to the durations of benthic macroinvertebrate life cycles in lotic systems (Turner and Dale, 1998; Snyder and Johnson, 2006). Such events
have been shown to be responsible for reductions in both benthic macroinvertebrate richness and density, as macroinvertebrates are not typically equipped to survive these catastrophic disturbances (Giller et al., 1991; Boulton and Lake, 1992; Lake, 2000; Synder and Johnson, 2006; Death, 2008). In Ireland, Giller et al. (1991) previously found that a catastrophic storm event (one in fifty year event) followed by several lesser flood events resulted in significant and long lasting impacts on the macroinvertebrate communities within the Araglin catchment in county Cork in the south of Ireland.

On 24 October 2011 a significant rainfall event (one in twenty to one in twenty five-year event: Mac Cárthaigh, 2011) occurred on the east coast of Ireland. This paper presents the results of an investigation of the impacts of that storm event on the benthic macroinvertebrate communities of four upland headwater streams in the Wicklow Mountains within the effected region. It also examines the ecological status of the four study streams before and after the storm and discusses the possible implications for status assessment and biodiversity of streams that have been exposed to such catastrophic storm events.

**METHODS**

Four headwater streams in county Wicklow were monitored over a four-month period from August 2011 to November 2011 (Fig. 1). The streams sampled included the Cransilliagh Brook (WM1), a tributary of the river Liffey (WM2), and the Ballyknocken (WM10) and Fraughan (WM11) brooks, both tributaries of Poulaphouca Reservoir (Fig. 1). The streams are either second or third order streams, with catchments ranging in size from 0.41 to 1.93 km², and the study sites were situated between 221 and 377 m in altitude (Tab. 1). All streams drain igneous (granite/felsite) geology and either blanket peat or poorly drained, peaty mineral soils (Tab. 1). Two streams drained semi-natural moorland (sites WM2 and WM11) areas and two drained plantation conifer forest (sites WM1 and WM10) (Tab. 1). The moorlands consisted of open landscapes generally dominated by purple moor grass (*Molinia caerulea* (L.) Moench), and heather (*Calluna vulgaris* (L.) Hull), with riparian banks dominated mainly by rushes (*e.g.* *Juncus* spp.) and gorse (*Ulex europaeus* L.). The forested streams drained catchments with approximately 30-35% conifer forest cover that did not create any significant shading. All streams sampled are low conductivity waters (<200 μS/cm).

In August 2011 each study stream was equipped with a WTW field continuous pH probe and a submerged pressure recorder (*diver* from Schlumberger Water Services, www.slb.com). WM1 was also equipped with a terrestrial *diver* to measure variations in atmospheric pressure. Changes in stream water levels were measured by subtracting the difference in pressure between the atmosphere

![Fig. 1. Location of the four study streams in the Wicklow Mountains on the east coast of Ireland. Wm1, Cransilliagh brook; Wm2, tributary of the Liffey river; Wm10, Ballyknocken brook; Wm11, Fraughan brook.](image-url)
Impact of a catastrophic storm on macroinvertebrates in headwater streams

and the submerged diver at each study stream and converting to an equivalent depth of water. Both devices recorded data at 5-min intervals. These data were collected during the catastrophic storm of 24 October 2011. However, some data were lost as some devices were damaged by the flood and some were discarded, e.g., the pH and water level data from streams WM1 and WM10, and the pH data from stream WM2 were deemed unreliable. The retrievable data from WM1, WM2 and WM10 were relatively similar to that of WM11 preceding the major storm event and therefore, only the full data set (i.e., both water level and pH) for WM11 is shown below. The rainfall data were sourced from Casement Aerodrome in south country Dublin, approximately 25 to 30 km north of sampling locations and were provided by Met Éireann (www.met.ie). The physical disturbance caused to each stream during the storm event was not measured quantitatively. However, the disturbance was visually assessed using before and after photographs (Fig. 2). For consistency with the water level and pH data only WM11 is shown below. Both photographs were taken facing upstream.

Macroinvertebrates were sampled at each site on two occasions, the first on 10 August 2011 before the storm

![Fig. 2. Upstream photographs of the study stream before (a) and after (b) the catastrophic storm event at Fraughan Brook (WM11) in the Wicklow Mountains.](image-url)

| Stream name      | Site code | Longitude   | Latitude   | Order | Elevation (m asl) | Catchment size (km²) | Stream width (m) | Mean catchment slope (degrees) | Soil | Geology | Land use |
|------------------|-----------|-------------|------------|-------|-------------------|----------------------|------------------|--------------------------------|------|---------|----------|
| Cransilliagh Brook | WM1       | 6° 22' 15.82" W | 53° 10' 4.23" N | 2     | 339               | 0.41                 | 2                | 9.72                           | BktPt | Gr/F    | Forest   |
| Tributary of river Liffey | WM2       | 6° 19' 59.40" W | 53° 9' 4.83" N | 3     | 377               | 1.93                 | 3                | 5.01                           | BktPt | Gr/F    | Moor     |
| Ballyknocken Brook | WM10      | 6° 30' 15.92" W | 53° 5' 52.10" N | 2     | 221               | 1.13                 | 2                | 12.0                           | AminSRPT | Gr/F    | Forest   |
| Fraughan Brook    | WM11      | 6° 28' 44.84" W | 53° 6' 55.48" N | 2     | 227               | 0.84                 | 3                | 11.2                           | AminSRPT | Gr/F    | Moor     |

*Approximate width at sampling point; m asl, metre above sea level; forest, conifer forest plantation; moor, moorland; BktPt, blanket peat; aminSRPT, peaty lithosolic-podzolic soils; Gr/F, granite/felsite.
and the second on 8 November 2011 after the storm, using six replicate Surber samples (area 0.09 m², mesh size 250 μm), three in the margins and three in the riffle habitats, within each study stream. The positioning of the Surber sampler was selected using random number tables. All samples were preserved using 70% industrial methylated spirits. In the laboratory, each sample was sorted after being washed through a 250 μm sieve. All samples sorted were independently assessed as part of routine quality control (i.e. full removal of all individual taxa) (Haase et al., 2006; Feeley et al., 2012). The Ephemeroptera, Plecoptera and Trichoptera were identified to genus-level, while Diptera, Coleoptera, Odonata, Mollusca and Hemiptera were identified to family level using standard Freshwater Biological Association identification keys. All other taxa were identified to the lowest feasible taxonomic level. Standardised taxon lists were produced and used to calculate mean taxonomic richness and densities, and mean richness and density of the Plecoptera, Ephemeroptera, Coleoptera and Diptera. These data were then tested for normality using Kolmogorov-Smirnov tests and pre and post storm values were compared using non-parametric Wilcoxon signed rank tests in PASW Statistics 18 (IBM SPSS Inc. 2010). The overall change in macroinvertebrate composition and structure was visualised using Cluster analysis dendograms based on Bray-Curtis similarities of untransformed data (PRIMER 6.1.12: Clarke and Gorley, 2006). Similarity profiles (SIMPROF) tests (PRIMER 6.1.12: Clarke and Gorley, 2006) were used to indicate significant a priori unstructured hierarchical groupings (P<0.05) of similar macroinvertebrate community composition and structure between streams on both sampling dates (i.e. before and after the catastrophic storm event).

Finally, the effect of the catastrophic storm was examined using macroinvertebrate metrics which assess the ecological status of each site based on the scoring of sensitive taxa at family level. The three metrics calculated were the Small Stream Risk Score (SSRS) (Anonymous, 2005; Kavanagh et al., 2006), the Biological Monitoring Working Party (BMWP) Score and the Average Score Per Taxon (ASPT) (Hawkes, 1997), previously used in headwater stream assessment in Ireland (Callanan et al., 2008; Feeley et al., 2012). The before and after storm metric data were then tested for normality using Kolmogorov-Smirnov tests and compared using paired t-tests in PASW Statistics 18 (IBM SPSS, 2010).

RESULTS

Rainfall, water level, pH and physical disturbance

Daily rainfall amounts at Casement Aerodrome from August to early November 2011 ranged from 0 to 27 mm, with a mean of 2 mm day⁻¹ falling during the study period (Fig. 3a). The exception to this was the rainfall event of 24 October when 83 mm fell in a 24-h period (Fig. 3a). During this catastrophic event the majority of rainfall (65.8 mm) fell over a 4-h period (15.00-19.00 h), with 36.1 mm falling between 15.00 and 19.00 h and a 60-min maximum of 18.2 mm falling between 16.00 and 17.00 h. Three other notable total daily rainfall events occurred preceding the catastrophic event, the first and second on 30 September and the second on 8 November 2011 after the storm.
(23 mm) and 1 October (27 mm) respectively, and the third on 23 October (19 mm) immediately prior the catastrophic rainfall event of 24 October (Fig. 3a).

Over the course of the study, stream water levels mirrored rainfall, with notable increases occurring when daily rainfall amounts were >20 mm (Fig. 3b). During the study period the most substantial change in water level occurred between 23 and 25 October, with the highest water level change occurring on 24 October (Fig. 3b). Water levels began to rise on 23 October increasing rapidly by 102.8 cm before dropping back to 54.9 cm and then rising again around midday on 24 October, with the peak of the flood occurring at approximately 17.00 h (194.2 cm) before slowly receding to circum-normal levels (<40 cm) shortly after 02.00 h on 25 October (Fig. 3b). The peak water level was more than 450% that of mean normal stream water levels (~35 cm) preceding the catastrophic event (Fig. 3b). Similar results were found at WM2, with a peak high water level of 128.2 cm at 19.50 h, >500% the mean water level (20.9 cm) (Fig. 3b). No reliable data were recorded at WM1 and WM10 due to probe damage.

The baseflow pH of all four study streams was generally circumneutral (~6 to 7 pH), and decreased with increases in rainfall and stream water levels. The changes in pH in stream WM11 highlights the episodic acidic nature of the stream in response to the increased rainfall and water levels, with a minimum pH (≤ 4.5 pH) recorded on three occasions during this study period (Fig. 3c). These occurred on 1 October, and again in quick succession on 23 and 24 October during the catastrophic storm event (Fig. 3c). The lowest pH (4.35 pH) was recorded at 15.25 h on 24 October (Fig. 3c). Unfortunately, no pH data could be reliably interpreted from WM1, while no pH data were received from WM2 and WM10 due to loss of equipment.

The instream characteristics of all four study streams before the storm event on 24 October reflected natural, stabilised riffle-run sequences, interspersed with shallow pools (Fig. 2a). All four streams were generally dominated by cobbles with some coarse gravel and sand, interspersed with a few large boulders (Fig. 2a). Instream vegetation consisted of algae, mosses and grasses. However, the post storm photograph indicates the magnitude of the instream disturbance caused by the high rainfall and increased water levels (Fig. 2b). The streambed experienced downcutting and extreme scouring, removal of most instream vegetation and an influx of boulders and coarse substrates (Fig. 2b). Furthermore, the stream banks were extremely eroded and destabilised (Fig. 2b), a characteristic noted across all four streams, with two streams (WM1 and WM10) altering their course for short distances (~2 to 5 m) at several points along their course.

**Effects on macroinvertebrates**

Taxon richness of the four study sites changed significantly (P<0.001) between August and November 2011 (Tab. 2). The total density of macroinvertebrates was also significantly reduced (P<0.001) on average by over 85% between the two sampling dates with average densities dropping from 2195 individuals m⁻² to 323 individuals m⁻² (Tab. 2). Changes in specific macroinvertebrate groups were more variable. Both plecopteran richness and density were not significantly (P>0.05) altered, with richness increasing slightly from 3 to 3.5 taxa 0.09 m⁻² even though densities dropped by over 53% on average (Tab. 2). Similarly ephemeropteran richness did not significantly change (P>0.05). However, ephemeropteran density across the four

| Scores                                      | Pre-storm (mean±SE) | Post-storm (mean±SE) | Z     | P value     |
|---------------------------------------------|---------------------|----------------------|-------|-------------|
| Taxon richness°                            | 13.7 (±0.5)         | 8.7 (±0.8)           | -4.054| <0.001*     |
| Total density#                             | 2195.0 (±242.0)     | 323.3 (±57.2)        | -4.286| <0.001*     |
| Plecoptera richness°                       | 3.0 (±0.3)          | 3.5 (±0.3)           | -1.285| =0.199      |
| Plecoptera density#                        | 327.9 (±102.9)      | 153.5 (±39.0)        | -1.815| =0.070      |
| Ephemeroptera richness°                    | 1.1 (±0.7)          | 0.9 (±0.2)           | -0.814| =0.415      |
| Ephemeroptera density°                     | 482.4 (±85.8)       | 42.5 (±11.0)         | -4.286| <0.001*     |
| Trichoptera richness°                      | 3.1 (±0.2)          | 1.7 (±0.3)           | -3.672| <0.001*     |
| Trichoptera density°                       | 227.5 (±41.1)       | 49.9 (±13.2)         | -4.287| <0.001*     |
| Coleoptera richness°                       | 2.3 (±0.2)          | 0.6 (±0.2)           | -3.672| <0.001*     |
| Coleoptera density°                        | 164.2 (±39.8)       | 15.7 (±5.1)          | -4.247| <0.001*     |
| Diptera richness°                          | 2.9 (±0.7)          | 1.4 (±0.2)           | -3.993| <0.001*     |
| Diptera density°                           | 918.1 (±153.3)      | 56.0 (±12.5)         | -4.286| <0.001*     |

°Taxa 0.09 m⁻²; †individuals m⁻²; Ephemeroptera, Plecoptera and Trichoptera, genus level identification; Coleoptera and Diptera, family level identification; *significantly different.
study streams dropped from an average of 482.4 individuals m$^{-2}$ to less than 43 individuals m$^{-2}$, a 91% reduction (Tab. 2). Trichopteran, coleopteran, and dipteran richness and density also significantly (P<0.001) dropped after the catastrophic storm event (Tab. 2).

Multivariate cluster analysis highlighted the similarities in benthic macroinvertebrate community composition and structure across study streams, with a >60% similarity pre-catastrophic storm event (Fig. 4). After the catastrophic storm event, similarities in macroinvertebrate structure between streams dropped below 50% (Fig. 4). Similarity profile tests highlighted significant differences between sites before and after the storm event on 24 October (P>0.05: SIMPROF), indicating a significant change in community composition and structure in all four study streams (Fig. 4.).

The mean SSRS and BMWP scores showed small decreases in site values, while the mean ASPT scores increased slightly (Tab. 3.) However, this resulted in no significant change in the ecological status of the sites following the storm event (Tab. 3).

**DISCUSSION**

The total daily rainfall at Casement Aerodrome on 24 October 2011 was calculated to be a one in twenty to one in twenty-five year event, with the four hour period between 15.00 and 19.00 h equalled a one in eighty year event (Mac Cáithigh, 2011). The hourly maximum of 23 mm recorded between 16.00 and 17.00 h was the highest hourly rainfall ever record for Casement Aerodrome (Mac Cáithigh, 2011). Although the exact amounts of rainfall in the Wicklow Mountains at the time are difficult to calculate, they are likely to be greater than that at Casement Aerodrome (92 m asl) because of the effect of altitude (Yang, 2004), although the temporal pattern is most likely very similar.

![Fig. 4](Image)

**Tab. 3.** Paired $t$-tests comparing the mean (±SE) of selected bioassessment metric scores before (pre) and after (post) the catastrophic storm event on the 24th of October 2011 in the Wicklow Mountains.

| Bioassessment metric | Pre-storm (mean±SE) | Post-storm (mean±SE) | $t$ value ($df=3$) | P value |
|----------------------|---------------------|----------------------|--------------------|---------|
| SSRS                 | 8.2 (±0.7)          | 8.0 (±0.9)           | 0.397              | 0.718   |
| BMWP                 | 95.0 (±6.7)         | 91.5 (±16.1)         | 0.291              | 0.790   |
| ASPT                 | 6.0 (±0.2)          | 6.5 (±0.3)           | -1.190             | 0.320   |

SSRS, Small Stream Risk Score; BMWP, Biological Monitoring Working Party; ASPT, Average Score Per Taxon.
This is further supported by the water level data recorded in the Fraughan Brook (WM11) with maximum levels recorded at 17.00 h, mirroring the period of highest rainfall recorded at Casement. The associated acidity (pH—4.5) was typical of episodic events known to occur in the Wicklow Mountains (Kelly-Quinn et al., 1996, 1997, 2008).

The impact of this catastrophic event on the benthic macroinvertebrate communities of the four study streams was similar to findings elsewhere, with the severe loss of macroinvertebrate richness and densities across all functional groups (Giller et al., 1991; Boulton and Lake, 1992; Angradi, 1997; Lake, 2000). During this study the pre-storm community structure across the sites was typical of episodic acid streams in the Wicklow Mountains (Tierney et al., 1998; Kelly-Quinn et al., 2008; Feeley et al., 2011, 2012). The significant loss of macroinvertebrate taxa during the storm probably resulted from the combined effect of high water level (and associated high flow), low pH, abrasion and movement of instream substrates and associated physical disturbance. The toxic effects associated with acidic conditions cause stress on sensitive macroinvertebrate biota, resulting in the drifting of taxa (e.g. Ephemeroptera) which is known to have a detrimental effect on population composition and structure (Ormerod et al., 1989; Kowalik and Ormerod, 2006). However, physical disturbance and wash-out of biota are likely to have played a more significant role in the changes captured during this study. Physical upheaval and disturbance resulting from high flow causes the streambed sediment to break up, allowing entrainment (i.e. the picking up and setting into motion) and abrasion of surface and subsurface substrates within and adjacent to the stream (Andrews, 1984; Andrews and Nankervis, 1995). Prior to the storm event on 24 October the instream habitats of the four sites were characterised by short riffle-run sequences, interspersed with shallow pools and were generally dominated by cobbles with some coarse gravel, sand and scattered large boulders, which is typical of small upland headwater streams in Ireland (Callanan et al., 2008; Feeley et al., 2012). The entrainment disturbance caused during the catastrophic event altered the natural sequence of instream habitats, causing severe damage to both the streambeds and banks in the four streams with the substrate composition changing to predominantly boulder and sand. This would have decreased the overall habitat heterogeneity. Instream vegetation was also visually reduced. A previous study by Englund (1991) showed that such stream disturbances have detrimental effects on stream moss, reducing the biomass and affecting species composition. In addition, during this study many benthic macroinvertebrate groups showed significant reductions in densities most likely as a result of the physical disturbance, since acid-tolerant benthic macroinvertebrate taxa, such as Siphlonurus, Ameletus and Plectronemia were also severely affected and ephemeropteron (acid-sensitive) richness was not notably reduced. The physical disturbance almost certainly caused extensive shifts in patches of unstable/distabilised sediment during the increased discharge, with large-scale invertebrate drift/loss across the community. Macroinvertebrate loss has been strongly linked to substrate movement as individuals shift out of the rising channel and are washed away and/or crushed by moving substrate and debris (Brittain and Eikeland, 1988; Death, 2008; Gibbons et al., 2010).

The combination of factors associated with the storm event altered the overall community composition and structure significantly. The Plecoptera were the only macroinvertebrate group to escape major change, with richness increasing slightly, although densities suffered losses of 54% in overall numbers. Similarly, ephemeropteron richness also remained unaffected, while densities were significantly reduced. The Trichoptera, Coleoptera and Diptera also suffered significant reductions in both richness and density within all four streams sampled. Although, some of the changes in composition are likely to be seasonal, with several taxa (e.g. Siphlonurus sp.) emerging as adults in mid to late autumn and other taxa (e.g. Ameletus sp.) having hatched during the same period, the fact the taxonomic resolution was predominantly at genus and family level indicates a severe level of impact on the overall community. The impact recorded would likely have been greater if species level identification had been utilised for all macroinvertebrate groups. Furthermore, the period of September through November is the principal recruitment/egg hatching period of many benthic macroinvertebrates in Ireland and elsewhere, especially the Plecoptera, Ephemeroptera and Trichoptera (Hynes, 1977; Crichton et al., 1978; Edington and Hildrew, 1995; Smith et al., 2000; Feeley et al., 2009; Elliott and Humphesch, 2010; McCarty, 2010). Although this was observed for plecopteran richness, the richness and density of the other macroinvertebrate groups should have, in theory, increased, or at least been maintained, rather than decreased during the study period.

The greatest heterogeneity of macroinvertebrate communities in riverine systems in Ireland generally occurs within the headwaters (Callanan, 2009; Feeley and Kelly-Quinn, 2012). The effects of large disturbances, such as massive storm events, vary from species to species and from habitat to habitat leading to higher heterogeneity within affected stream communities (Reice, 1985; Resh et al., 1988; Townsend et al., 1997; Schwendel et al., 2011b). A decrease in percentage similarity in community composition and structure among the sampled sites was seen post storm, indicating an increase in taxonomic heterogeneity amongst the affected sites. Although not directly compared, the impact on macroinvertebrate communities did not vary with land use. Previous research
throughout North Western Europe found little difference in peak flows recorded in moorlands and forested streams (Robinson et al., 2003).

An awareness of the effects of such massive storms is important when one is determining ecological status. The three bioassessment metrics used during this study showed no overall change in the ecological status of the study sites post storm. This most likely relates to the lack of any organic pollution and the maintenance of both Plecoptera and Ephemeroptera richness which helps preserve the metric values. This is an important finding and suggests that these metrics, at least in clean water systems, are unaffected by significant flooding. The catastrophic losses of macroinvertebrate densities demonstrate that there clearly was significant alteration to the macroinvertebrate composition and structure of examined streams. It is, therefore, vital that metrics are developed or altered to take account of dramatic losses in densities/abundances of taxa, such as that developed by Schwendel et al. (2011b) for assessing streambed stability in New Zealand’s North Island.

Catastrophic flooding of the type described in this study has potential implications for site recovery and thus, regional biodiversity, particularly if such events are as spatially widespread as this storm (Mac Cárthaigh, 2011) and were to become a more common occurrence as predicted due to climate change (see McElwain and Sweeney, 2007; Sweeney et al., 2008). Such catastrophic events with major depletions within a system(s) are likely to affect the resource pool for recolonisation and in some instances may actually result in species exhaustion (Giller et al., 1991; Gibbons et al., 2010). If frequent physical disturbances occur, it possibly limits benthic macroinvertebrate recovery, both directly through continuous/frequent streambed destabilisation, reduced habitat heterogeneity, loss of critical life cycle stages and refugia and/or competitive exclusion, or indirectly through the alteration of the food resource and vegetation structure of the stream (Townsend et al., 1997; Robinson and Minshall, 1986; Tixier et al., 2009; Schwendel et al., 2011a).

Therefore, the significant disturbance, and loss of both taxonomic richness and density, if recurrent, will have major implications for the maintenance of healthy macroinvertebrates populations, and thus ecological diversity within affected headwaters. The downstream persistence and effects of such large storms on the surface water hydrology and ecology are inextricable related to the changes in flow and magnitude of bedrock/sediment scouring, entrainment and transport, which vary with catchment size (Richards et al., 2001; Jakob et al., 2003). Consequently, the ability of any taxonomic groups within affected watercourses to resist the effects of a major disturbance depends on several biotic and abiotic factors. Such factors include the stability of instream habitat, the attachment/burrowing ability of taxa and the effective use of refugia and flexible/plastic life history strategies (Giller et al., 1991; Palmer et al., 1996; Tixier et al., 2009).

CONCLUSIONS

Catastrophic storm events in upland headwater streams can severely alter macroinvertebrate composition and structure, with major losses in the densities of all taxonomic groups. The resultant losses may have long term implications as many studies have indicated the slow and often poor recovery of macroinvertebrate communities (Giller et al., 1991; Gibbons et al., 2010). Regardless, the recovery of the natural taxonomic richness, density, and community composition and structure post major disturbance is a function of the magnitude of any impact and is therefore likely to be both time and site specific (Wallace, 1990; Death, 2008). For example, Giller et al. (1991) found season to be a factor, with winter invertebrates much more resilient to disturbance compared to summer taxa due to life cycle disruption of summer taxa and the adaptability of winter taxa. However, any recovery of benthic macroinvertebrate groups is reliant on several factors such as the utilisation of instream refugia, survival in the drift (i.e. safe resettlement), life cycle dynamics and recolonisation sources (Giller et al., 1991; Palmer et al., 1996; Tixier et al., 2009).

Ultimately, such massive impacts and the potential slow recovery may have major implications for future stream diversity. Several studies highlight the importance of individual small headwater streams to regional diversity in Ireland, the UK and the US (Minshall et al., 1985; Minshall and Robinson, 1998; Furse, 2000; Callanan, 2009; Feeley and Kelly-Quinn, 2012). Given the spatial influence of this storm event (Mac Cárthaigh, 2011), the level of impact across all four study sites, and the high number of small headwater streams in the Wicklow Mountains, the potential impact as a result of such catastrophic storms on the regional macroinvertebrate populations is high, possibly reducing local recolonisation sources and thus, conceivably limiting the ability of affected streams to maintain at least good quality waters as prescribed by the Water Framework Directive (European Parliament and Council, 2000).

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