Bioaccumulation of organochlorine pesticides (OCPs) in molluscs and fish at the Sai Gon - Dong Nai estuary

Tích lũy sinh học hóa chất báo vẽ thực vật clo hữu cơ trong nguyên thể và cá tại cầu sông Sài Gòn - Đồng Nai

Research article

Nguyen Xuan Tong1,2, Tran Thi Thu Huong3, Duong Thi Thuy1,4*, Mai Huong5, Duong Trong Khang2, Huynh Cong Luc6, Pham Thi Loan7, Le Thi Phuong Quynh

1Graduate University of Science and Technology, Viet Nam Academy of Science and Technology (VAST); 2Institute for Environmental Science, Engineering and Management, Industrial University of Ho Chi Minh City; 3Faculty of Environmental, Ha Noi University of Mining and Geology; 4Institute of Environmental Technology, VAST; 5University of Science and Technology of Ha Noi, VAST; 6Institute of Natural Product Chemistry, VAST, 18 Hoang Quoc Viet, Cau Giay, Ha Noi, Vietnam

The aim of this study is to assess the biological accumulation of pesticide residues in aquatic organisms in Sai Gon - Dong Nai (SG-DN) estuary. Fish and mollusks were collected directly at the Soai Rap and Long Tau estuary of the SG-DN river system, washed and separated for taking the tissue. The organochlorine compounds from the tissue were then extracted and analyzed by gas chromatography system. The results showed that, the concentration of OCPs in Tegillarca granosa, Mercenaria auricularia and Bostrychus sinensis varied from 6.4 to 59.9 µg/kg, 7.2 to 322 µg/kg, and 2.9 to 114.3 µg/kg, respectively. In general, molluscs species that accumulate more heptachlor, aldrin, endrin or dieldrin tend to accumulate less DDT (dichlorodiphenyltrichloroethane). Endosulfan was the most commonly found in three bivalve mollusks while DDTs (1.5–75.2 µg/kg, averaging 8.7 µg/kg) was the most popular OCPs in the fish (Bostrychus sinensis) samples. In DDT group, the p,p'-DDT metabolite accounted for the largest percentage, reaching 50% of total DDTs. In HCH (Hexachlorocyclohexane) group, β-HCH isomer was predominant in almost samples.

Keywords: biological accumulation, estuarine, fish, molluscs, pesticide

* Corresponding author
Email: duongthuy0712@gmail.com
1. Introduction

The Sai Gon – Don Nai (SG-DN) River system with a large estuarine area is the main water sources for agriculture and the daily activities of people living along the river banks in Ho Chi Minh City and the neighbouring provinces (Ba Ria Vung Tau, Dong Nai, Long An and Binh Duong). However, rapid development in this region raised concerns about local environment and ecological integrity (Anh, 2003; Nguyen et al, 2007). Large amounts of untreated municipal and industrial wastewater, as well as runoff from landfills carrying numerous toxic complex contaminants are released directly into the river. The increase in population and productive activities, as well as use of pesticides without logical in agricultural production increased pesticide residues in the river. Since many years, this estuary has been served as one of the main areas for large-scale aquaculture. Many species of fishes and molluscs such as oysters, clam etc. were cultivated in here. These species have a long growing cycle, ranging from 8 to 10 months and need to live in a clean environment without contamination. Therefore, the protection of river water quality is an important task for sustainable development, especially when demand for water supply and use of fishery products increases rapidly.

Organochlorine pesticides (OCPs) which are large groups of crop protection chemicals have been widely used for various applications during the past several decades. Due to their persistent, semi-volatil, toxic properties and highly bio-accumulation especially in lipophilic substances and tissues, organochlorines distribute ubiquitously in the global environment and can be detected at all levels of food chains, including humans. Many OCPs are classified into persistent organic pollutants (POPs), according to Stockholm formula (UNEP, 2011). These compounds affect seriously public health of humans and wildlife because they alter normal functions of endocrine (endocrine disruptor pesticides) (Mnif et al, 2011), interfere with female hormonal function, leading to negative effects on the reproductive systems (Bretveld et al, 2006), have the close relationship with neurodegeneration diseases (Dana, 2008), many popular cancers such as breast cancer (Wolff et al, 2007; Bhatnagar et al, 2002), non Hodgkin’s lymphoma (Cantor et al, 1992), and other types of cancer (Alavanja et al 2013).

Although the use of many organochlorine compounds such as Lindane, DDT, Endrin, Dieldrin etc. has been prohibited according to Circular No. 21/2013/MARD because of the long half-life and difficulty for controlling persistent residue (MARD, 2013). However, pesticide contamination still occurs and is difficult to handle, including the issue of OCPs bioaccumulation in the environment. Some reports published recently focus on the assessment of OCPs and heavy metal pollution in sediments, rivers, seas and estuaries along Vietnam such as: Mekong River Delta (Nguyen et al, 2006), Sai Gon - Dong Nai River (Nguyen et al, 2007), Tien estuary (Nguyen et al, 2017), Thi Vai Estuary and Can Gio Mangrove Forest (Costa-Böddeker, 2016), Nha Trang Bay (Hoang et al, 2015), Red River Delta and along the coast of North Vietnam (Dang et al, 1998). Until now, there is a few of studies investigating the toxic OCPs bioaccumulation in Tegillarca granosa, Meretrix lyrata, Margaritifera auricularia and Bostrychus sinensis cultured in the SG-DN estuary. Therefore, the aim of this work is to investigate levels of OCPs residues and their isomer including of the groups such as: HCHs, DDTs, Aldrin, Dieldrin, Heptachlor, Endosulfans in these species in SG-DN estuary area.

2. Material and methods

2.1. Study area

The current study was conducted in the Long Tau and Soai Rap estuary belonging to the Sai Gon-Dong Nai estuary, located in Ho Chi Minh city, Vietnam (Figure 1). The estuary with funnel shape has a total catchment area of 43,500 km², including areas contiguous with the Ba Ria - Vung Tau, Dong Nai, Long An and Binh Duong provinces. The river catchment has two distinct seasons: the dry season is from December to April and the rainy season is from May to November, with total annual rainfall from 170.8 – 212.4 mm. There are four main rivers (Soai Rap, Long Tau, Thi Vai and Cai Mep) collective adding water to the SG-DN River system, before coming to the East Sea with the total annual discharge of approximately 34 billion m³ (Le et al, 2015). Currently, the area of the paddy field in the catchment of the SG-DN river system is around 18,000 hectares.

![Figure 1. The SG-DN estuary and sampling areas](image)

2.2. Sampling

2.2.1. Sample identification

For sample identification, all the samples are classified according to their morphological characteristics and explicit enumeration to the David's category (1969): scallops,
clams, mussel and fish. The bivalve molluscs (clams, oysters and clams) are brushed with a brush, washed with water to remove the adhering detritus.

2.2.2. Sampling

Thirty-eight sites surrounding the boundary of the Long Tau and Soai Rap estuary belong to the SG-DN system were selected for the current study (Figure 1). Samplings were conducted in two seasons (rainy season: May 2017, dry season: October and November 2017). The sampling process is similar to the procedure of two authors Kožul (2008) and Mohamed (2012). The experimental samples including: scallops, mussel, clams and fish were taken directly and randomly by hand. All the sample were immediately kept in plastic boxes and stored in an ice-chest at 5°C and transported to the laboratory for analysis.

2.3. Procedure for sample treatment and analysis

The processing analysis is followed the steps presented in the Figure 2 (Kožul, 2008; Shreadah, 2012).

\[ 20g + Na_2SO_4 \]
\[ \downarrow \]
\[ \text{To extract by ultrasonic and centrifuge 3 times by acetone} \]
\[ \downarrow \]
\[ \text{Aqueous extract layer were transferred n-hexane buffer} \]
\[ \downarrow \]
\[ \text{Rotation vacuum machine to collect 5ml extract} \]
\[ \downarrow \]
\[ \text{A solid-phase column containing 8 grams of activated Florisil} \]
\[ \downarrow \]
\[ \text{Rinse by two phase} \]
\[ \text{Phase 1: by n-hexane for analysis PCBs} \]
\[ \text{Phase 2: by n-hexane:DCM (4:1,v/v) for analysis OCPs} \]
\[ \downarrow \]
\[ \text{Rotation vacuum machine to collect 5ml extract} \]
\[ \downarrow \]
\[ \text{Transfer to test tube for removing lipid} \]
\[ \downarrow \]
\[ \text{Rotation vacuum machine by N}_2\text{ air to collect 1ml extract} \]
\[ \downarrow \]
\[ \text{Analysis OCPs by GC-ECD} \]

Figure 2. The procedure for analyzing OCPs in experimental samples

| Table 1. Organochlorine pesticides detected (µg/kg) in all sample |
|-----------------|-----------------|-----------------|-----------------|-----------------|
| Species         | Margaritifera auricularia (N = 8) | Tegillarca granosa (N = 13) | Meretrix lytara (N = 9) | Bostrychus sinensis (N = 8) |
| Sample size     | Total HCHs      | SEM             | Range           | Total HCHs      | SEM             | Range           | Total HCHs      | SEM             | Range           | Total HCHs      | SEM             | Range           |
|                 | 6.527           | 1.914           | 2.001 – 16.591 | 5.005           | 0.925           | 1.706 – 12.149 | 7.160           | 4.344           | 1.742 – 41.832 | 4.939           | 1.509           | 0.804 – 13.788 |
| α-HCH           | 1.759           | 0.492           | 0.745 – 4.128  | 0.717           | 0.288           | 0.745 – 4.128  | 1.515           | 0.905           | 0.0 – 3.695    | 0.970           | 0.367           | 0 – 3.132       |

2.4. Statistical analysis

All experiments were done in triplicate and the data were calculated as mean ± SD (standard deviation) and drawn by the software SigmaPlot. Statistical significance was accepted at a level of p < 0.05.

3. Results and discussion

3.1. The analysis results of OCPs component

In this study, the individual numbers of the analysed species M. auricularia, T. granosa, M. lyrata and B. sinensis were 8, 9, 13 and 8, respectively. Table 1 shows the average mean, the standard error of the mean and the varied values of OCPs in all analysed samples. According to Vietnamese standard, 12 different common components of OCPs and pesticides residues were detected with high contents in the both three bivalve molluscs and fish B. sinensis species collected in the study area (MARD, 2013; 2015). The bioaccumulation of OCPs in the survey species is significantly different. Almost all individuals of Bostrychus sinensis did not accumulate endosulfan II; many individuals of Tegillarca granosa did not accumulate α-HCH and δ-HCH; Meretrix lyrata individuals did not accumulate aldrin, sulphate endosulfan and Margaritifera auricularia did not accumulate aldrin.

The highest total OCP accumulation in M. lyrata was 114.317 µg/kg, while the lowest in M. auricularia varied from 4.472 to 62.114 µg/kg by wet weight. The analysis results showed that the highest total OCP accumulation in Meretrix lyrata was 114,317 µg / kg, while the lowest in Margaritifera auricularia varied from 4,472 to 62,114 µg/kg by wet weight. However, the endosulfan group was found in all three bivalve molluscs with the content of 1.5 - 75.2 µg/kg and the DDT group was the most common OCPs in the fish B. sinensis with the content of 8.7 µg/kg by wet weight. The high accumulation of HCHs in fish sample indicated the pollution in the environment and the risk to human health.
### 3.2. The analysis results of the organochlorine derivative groups

Figure 3 illustrates the content of seven types of pesticides affecting 4 studied species: *M. auricularia*, *T. granosa*, *M. lyrata*, and *B. sinensis*. In general, it can be seen that endosulfans were the most predominant compounds in three molluscs species; followed by DDT, HCH, heptachlor and the rest substances. However, the content of substances high or low is very different in different species. The contents of endosulfans in *M. lyrata*, *T. granosa*, *B. sinensis* and *M. auricularia* respectively were 27,831 µg/kg, 14,482 µg/kg, 9,674 µg/kg and 4,591 µg/kg, respectively. Followed by the content of DDT was 6,083 µg/kg in *T. granosa* and the lowest content in this species is the HCH group. In contrast, in fish *B. sinensis*, DDTs were the most predominant OCPs (16,662 µg/kg), followed by endosulfans and HCH derivatives. However, in *M. lyrata*, heptachlor had the relatively equal amount to HCHs and DDTs. 

| Species                     | Margaritifera auricularia (N = 8) | Tegillarca granosa (N = 13) | Meretrix lyrata (N = 9) | Bostrychus sinensis (N = 8) |
|-----------------------------|---------------------------------|-----------------------------|------------------------|----------------------------|
| Sample size                 | 3.107                           | 2.076                       | 3.330                  | 2.246                      |
| β-HCH                       | SEM 1.098                       | 0.482                       | 2.058                  | 0.783                      |
| Range                       | 0.817 – 10.012                  | 0 – 6.455                   | 0.586 – 19.704         | 0.221 – 5.458              |
| δ-HCH                       | SEM 0.686                       | 0.889                       | 0.389                  | 1.224                      |
| Range                       | 0.313                           | 0.445                       | 0.195                  | 0.483                      |
| γ-HCH                       | SEM 0.976                       | 1.323                       | 1.925                  | 0.500                      |
| Range                       | 0.245                           | 0.392                       | 1.247                  | 0.199                      |
| Total endosulfans           | 4.591                           | 14.482                      | 27.831                 | 9.674                      |
| SEM                         | 1.188                           | 2.952                       | 24.02                  | 15.836                     |
| Range                       | 0.534 – 9.862                   | 0.813 – 30.615              | 0.205 – 219.491        | 0 – 47.061                 |
| Endosulfan I                | SEM 2.405                       | 1.415                       | 1.059                  | 1.737                      |
| Range                       | 1.199                           | 0.661                       | 0.454                  | 0.7                        |
| Endosulfan II               | SEM 0.456                       | 2.620                       | 0.531                  | 0.164                      |
| Range                       | 0.13                            | 0.962                       | 0.164                  | 0.109                      |
| Endosulfan Sulfat           | 1.730                           | 10.446                      | 26.241                 | 7.772                      |
| SEM                         | 0.868                           | 3.098                       | 23.707                 | 5.637                      |
| Range                       | 0.105 – 5.743                   | 0 – 29.679                  | 0 – 215.414            | 0 – 46.507                 |
| Total OCPs                  | SEM 28.020                      | 36.992                      | 51.585                 | 35.750                     |
| Range                       | 7.105                           | 4.488                       | 34.026                 | 13.326                     |
| Heptachlor                  | 1.360                           | 3.516                       | 7.738                  | 1.003                      |
| SEM                         | 0.315                           | 0.755                       | 3.853                  | 0.256                      |
| Range                       | 0.178 – 2.917                   | 0 – 8.121                   | 0 – 30.524             | 0 – 2.056                  |
| Aldrin                      | 1.002                           | 1.713                       | 0.458                  | 1.245                      |
| SEM                         | 0.547                           | 0.474                       | 0.337                  | 0.43                       |
| Range                       | 0 – 4.380                      | 0 – 5.421                   | 0 – 3.055              | 0 – 3.815                  |
| Dieldrin                    | 4.626                           | 2.793                       | 0.201                  | 1.743                      |
| SEM                         | 3.441                           | 1.756                       | 0.074                  | 0.564                      |
| Range                       | 0.055 – 27.816                  | 0 – 23.157                  | 0 – 0.557              | 0 – 3.822                  |
| Endrin                      | 1.675                           | 3.359                       | 1.936                  | 0.589                      |
| SEM                         | 1.252                           | 1.317                       | 0.916                  | 0.217                      |
| Range                       | 0.039 – 10.391                  | 0 – 17.351                  | 0 – 8.341              | 0 – 1.580                  |

* SEM: Standard error of the mean
gillarca granosa was more predominant in (62%), followed by percentage, reaching 50% of DDTs total in all samples. In partic-
tion of total HCHs was observed in M. lyrata (1.241), followed by M. sinensis (9,673 µg/kg) and the lowest value of only 4.591 µg/kg was in M. auricularia. In figure 5, except M. auricularia samples, endosulfan sulphate tended to be more active than other endosulfan derivatives in the remaining three species. It accounted for the largest percentage, reaching up to 80-90% of endosulfans total and this value in M. lyrata (26,241 µg/kg), T. granosa (10,441 µg/kg) and B. sinensis is 7,772 µg/kg, respectively. For M. auricularia, total endosulfans was slightly lower than endosulfan I, this value is 1.73 µg/kg) compared with 2.4 µg/kg endosulfan I. According to Dang et al (2001), in the canals in Hanoi city, p,p'-DDE always accounts for the largest proportion, up to 70-80% of DDT in freshwater mussel (Angulyagra sp.).

The study of bioaccumulation OCPs in some species of bi-valve molluscs and fish in the SG-DN estuary showed that there was a difference in the level of bioaccumulation in each species. In general, molluscs species tended to accumu-
late heptachlor, aldrin, endrin or dieldrin more than DDTs. This result is different from the report of Dang et al (2001) and Marta (2015). (Commendatore et al, 2015) showed that organochlorine bioaccumulation processes through bottom sediment resuspension were found in surface sediment and bivalve molluscs (29.4-206.0 ng/g dry weight); whereas imposex incidence was only 15% in the gastropod Pareuthria plumbea. Moreover, the presence of p,p’-DDE and the p,p’-DDE ratio likely reflect the high
stability of the stock compounds and their isomer in the aquatic environment compared with other metabolites. The OCP compounds in bivalve molluscs showed different patterns. This result may be due to different input pathways such as OCPs being incorporated into organisms directly from the water column’s suspended particles while being transformed in sediments. According to these authors (Dang et al., 1998; Commendatore, 2015), the organochlorinated pollution was clearly related to the migration of contaminants and the ability to self-disintegrate pesticides by environmental factors such as: temperature, atmosphere, transport etc.

4. Conclusion

This study demonstrates the high accumulation levels of organochlorine pesticides in different species of bivalve molluscs cultured in the Sai Gon – Dong Nai estuary compared with Vietnam standard 10:2008/MONRE. Endosulfan was the most commonly found in three bivalve molluscs (4,591-27,831 µg/kg, average 14,144 µg/kg by wet weight) while DDTs (6,124-16,662 µg/kg, average 9,4565 µg/kg by wet weight) was the most popular OCPs in the fish (B. sinensis) samples. In DDT group, the p,p'-DDT metabolite accounted for the largest percentage, reaching 50% of total DDTs. In HCH (Hexachlorocyclohexane) group, β-HCH isomer was predominant in almost samples. However, in order to have more consistent and scientific assessments, it is necessary to combine with the study of OCPs deposit in water, sediment and other aquatic parameters to limit the potential for bioaccumulation of the pesticides in these species.

5. References

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