SnO$_2$-Containing Clinoptilolite as a Composite Photocatalyst for Dyes Removal from Wastewater under Solar Light

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Abstract: Due to their adsorbent, ion exchange and catalytic properties zeolites are suitable for a variety of applications. We report on the photocatalytic activity of a readily available and inexpensive natural zeolite clinoptilolite (Z) containing SnO$_2$ (Sn-Z). The Sn-Z samples with 3–15 wt. % of Sn were prepared using a precipitation–deposition method. Powder X-ray diffraction analysis showed that the zeolite structure was unaffected by the introduction of the Sn-phase. Diffuse reflectance UV/VIS spectra of the Sn-Z samples confirmed the presence of SnO$_2$ and X-Ray absorption spectroscopy analyses suggested that the SnO$_2$ particles mainly resided on the surface of the clinoptilolite, while ATR-FTIR analysis gave some clues that part of the SnO$_2$ phase was incorporated in the pores of the zeolite. The presence of SnO$_2$ in Sn-Z increased both adsorption capacity and photocatalytic performance which could be partially explained by higher surface area and partially with an increased negative potential of the surface. Adsorption and total degradation of methylene blue (MB) for the Sn-Z with the highest amount of Sn (15 wt.%) was about 30% and 45%, respectively, suggesting a synergetic effect between SnO$_2$ and the clinoptilolite lattice. Reusability tests showed that these catalysts present a promising material for water purification.

Keywords: SnO$_2$; zeolite; SnO$_2$-clinoptilolite composite; photocatalysis; solar light; methylene blue removal; wastewater treatment

1. Introduction

Dyes are widely used in the textile and food industry as well as paper and photo–electrochemical cells’ production. This brings a burden of organic dye contamination to the environment, exerting a harmful effect on humans and animals [1,2]. Therefore, various methods for their remediation have been studied, including physical, chemical, and biological methods [3]. In recent years, heterogeneous photocatalysis has attracted considerable attention as an alternative and environmentally friendly process for the degradation and discoloration of organic dyes. Generally, it can be defined as the acceleration of the production of free radicals in the presence of a photocatalyst by its excitation with...
visible, ultraviolet, or infrared irradiation. Metal oxide semiconductors, such as titania (TiO$_2$), zinc oxide (ZnO), as well as tin oxide (SnO$_2$), have been used as effective photocatalysts [4–6].

SnO$_2$ has received considerable attention for fundamental research and industrial application due to its properties, such as low electrical resistance, high electrical conductivity, and high optical transparency in the visible region [7,8]. As an n-type semiconductor having a bandgap of about 3.6 eV at bulk state, SnO$_2$ can be a good alternative photocatalyst, but its photocatalytic ability for the degradation of pollutants is directly related to the synthesis procedure (morphology, particle size, and dopant concentrations). SnO$_2$ nanoparticles, synthesized by the microwave heating method were evaluated for the degradation of methyl violet 6B (MV6B) and methylene blue (MB) dye under direct sunlight. The synthesized SnO$_2$ nanoparticles showed excellent dye-sensitized visible light photocatalytic activity for the degradation of both MV6B and MB [9]. In addition, it is reported that SnO$_2$ prepared by the hydrothermal method shows the photocatalytic ability for the degradation of MB dye under UV light irradiation [10]. Recently, the photocatalytic activity of SnO$_2$ nanotubes fabricated by template-based liquid phase deposition and calcined at different temperatures was studied. The activity was tested in the degradation of MB under UV light irradiation, produced by four 100 W Xe lamps [11]. The reported results clearly indicated that photocatalytic activity increased with increasing calcination temperature due to an increase in specific surface area as well as their increased crystallinity.

Since the high specific surface area is one of the most important treats of a good catalyst, many materials have been used as catalyst supports. Zeolites have attracted great attention due to their adsorbent, ion exchange, and catalytic properties. Increasing interest has been applied toward natural zeolites, especially to clinoptilolite since it is the most abundant natural zeolite, inexpensive, and environmentally friendly. The porous structure of clinoptilolite is characterized by two systems of channels: the first one composed of eight- and ten-membered rings formed by linkage of AlO$_4$ and SiO$_4$ tetrahedral units, and the second one formed by eight-membered rings. These two pore systems are interconnected within the lattice hosting the extraframework/exchangeable cations and water molecules [12]. The possibility of chemical modification of the clinoptilolite structure allows it to be a prospective candidate to be grafted and further used as a catalyst. For example, CuO incorporated into nanoparticles of clinoptilolite was tested in the photo-discoloration process of an aqueous mixture of MB and bromophenol blue by irradiation under an atmospheric condition with a medium pressure Hg lamp–75 W [13]. The results demonstrated the importance of the clinoptilolite support to enhance the reactivity of CuO particles by preventing aggregation of CuO particles and also electron-hole recombination. Similarly, it has been reported for ZnO-supported clinoptilolite in photo-discoloration of bromothymol blue (BTB) aqueous under Hg lamp irradiation [14]. The results of SnO$_2$-containing clinoptilolite for photocatalytic usage are still scarce. However, coupling ZnO-SnO$_2$ particles onto clinoptilolite nanoparticles was investigated in the photocatalytic degradation of an aqueous mixture of 4-methylbenzoic acid and 2-chloro-5-nitrobenzoic acid under UV irradiation [15], as well as metronidazole photodegradation irradiated by moderate pressure Hg-lamp [16]. The latter was also studied by the coupling of SnO$_2$ and NiO and supporting them onto clinoptilolite nanoparticles [17].

Taking all this into account, the aim of this study was to investigate the photocatalytic activity of SnO$_2$-containing clinoptilolite in the photodegradation process of MB dye as a model cationic dye. It is a representative of reactive dyes, which generally consist of a chromophoric group (azo [-N=N-], anthraquinone, phthalocyanine, etc.), an anchor group (vinylsulfone, chlorotriazine, etc.) and a group which increases their water solubility. To date, coupling SnO$_2$ with natural clinoptilolite has only been scarcely reported as a catalytic system [16]. Moreover, most of the studies with SnO$_2$ as a photocatalyst were done under UV light [10,11,18–20]. Yet, for a commercial purpose, largely available visible light is a better choice. Consequently, in this study, photocatalytic experiments were performed by using illumination with visible-light lamps as a natural resource which reduces the cost of the process and environmental impact. We discuss the mechanism of action, SnO$_2$ content optimization, and long term use of such composites. Finally, since natural zeolite (clinoptilolite) is an
economically and environmentally friendly material, it is a promising component for the coupling with other photocatalysts as well.

2. Results

2.1. Characterization of the Materials

The X-ray diffraction (XRD) pattern of the synthesized pure SnO\(_2\) (Figure 1.) displayed characteristic diffractions at 2\(\theta\) values 26.6°, 33.9°, 38.1°, 51.9°, 61.9°, and 65.1° (database, Rutile-like structure, JPCDS 41-1445). The pattern of parent clinoptilolite (CLI) showed the presence of clinoptilolite, feldspar, and quartz. The prepared SnO\(_2\)-CLI composites showed the clinoptilolite crystallinity of the composite decreased with the amount of SnO\(_2\) and that additional diffraction peaks that could be ascribed to SnO\(_2\) did not appear. This can be explained by the fact that SnO\(_2\) particles at the CLI surface are amorphous. Similar results have been reported for clinoptilolite covered by Fe(III) and MnO\(_2\) particles [21,22].

![XRD patterns](Figure 1).

Figure 1. XRD patterns of SnO\(_2\) from database (SnO\(_2\)-Rutile), pure SnO\(_2\) (SnO\(_2\) synthesized), pure clinoptilolite (Clinoptilolite), and SnO\(_2\)-containing clinoptilolite with different SnO\(_2\) concentrations (SnO\(_2\)CLI05, SnO\(_2\)CLI15, SnO\(_2\)CLI25).

Once the phase composition was known, the samples were tested with thermogravimetric analysis (Figure 2) to see the possible evolution of SnO\(_2\) during heating. The weight loss up to 200 °C in the thermogram of the SnO\(_2\) precursor (Figure 2a) can be ascribed to the dehydration process, while the maximum in the DTG (differential of weight loss) curve, centered at 321 °C, can be attributed to the dehydration and dehydroxylation of amorphous tin hydroxide [23]. This is in accordance with the data reported for the formation of SnO\(_2\) by thermal decomposition of Sn(II) oxyhydrate under air [24]. XRD analysis confirmed the thermal stability of the composites up to 800 °C (data not shown). Thermal analysis (Figure 2b,c) shows that the total weight loss increased with the increase in Sn content from 12.3% (SnO\(_2\)CLI05) to 16.0% (SnO\(_2\)CLI25). DTG curves displayed a maximum at about 60 °C which can be ascribed to the loss of physically adsorbed water at temperatures below 100 °C. Furthermore, a broad maximum in the temperature range 200–400 °C and peaking at 350 °C was noticeable due to the dehydroxylation and formation of Sn oxide species in the clinoptilolite structure. The delay in the peak weight loss from 321 to 350 °C was noticeable and suggests the interaction of Sn species with the zeolitic support causing delayed dehydroxylation.
were O, Si, and Al, and, in minor amounts, Na, K, Ca, and Fe (Figure S1). The concentration of Sn onto SnO\textsubscript{2}-containing clinoptilolite increased with the increase in the initial concentration of the SnCl\textsubscript{2} solution used in the modification process demonstrating that Sn was loaded into the clinoptilolite structure. It is worth mentioning that the grafting efficiency was approximately 60% regardless of the nominal Sn amount (5, 15, or 25 wt. %) according to the ratio between measured Sn in composites and Sn put into the synthesis. In addition, it can be seen that the Si/Al ratio for pure clinoptilolite was 5.0, which is typical for this type of zeolite found in nature and in good agreement with other reports [25,26]. However, Si/Al ratios of SnO\textsubscript{2}-containing samples increased with respect to raw CLI. This can be ascribed to the partial dealumination process during the conversion of CLI to H–CLI [27,28]. This was also confirmed by the constant Si/Al ratios (~7.3) at higher Sn loadings.

Table 1. Average elemental composition (at. %) of clinoptilolite (CLI) and SnO\textsubscript{2}-containing clinoptilolite obtained by the energy-dispersive X-ray (EDX) analysis. For clarity, Sn contents are shown also as wt.%.

|       | CLI  | SnO\textsubscript{2}CLI\textsubscript{0.05} | SnO\textsubscript{2}CLI\textsubscript{0.15} | SnO\textsubscript{2}CLI\textsubscript{0.25} |
|-------|------|------------------------------------------|------------------------------------------|------------------------------------------|
| O     | 68.69| 70.2                                     | 71.57                                    | 72.89                                    |
| Mg    | 0.72 | -                                        | -                                        | -                                        |
| Al    | 4.7  | 3.52                                     | 3.14                                     | 2.91                                     |
| Si    | 23.68| 25.46                                    | 23.34                                    | 21.31                                    |
| Na    | /    | 0.11                                     | 0.08                                     | 0.04                                     |
| K     | 0.45 | 0.04                                     | 0.02                                     | 0.02                                     |
| Ca    | 1.34 | 0.04                                     | 0.06                                     | 0.06                                     |
| Fe    | 0.42 | 0.11                                     | 0.04                                     | 0.03                                     |
| Sn    | -    | 0.52                                     | 1.75                                     | 2.74                                     |
| Sn(wt.%) | - | 3.04                                     | 9.89                                     | 14.93                                    |
| Si/Al | 5    | 7.2                                      | 7.4                                      | 7.3                                      |

IOP-OES elemental analysis of CLI was also performed to complement the EDX elemental analysis; the results are shown in Table S1 (Supplementary Information), and they confirmed the findings of SEM-EDX method and found the values of elements in the zeolitic support to be concurring the fact that the support could be called clinoptilolite.

The presence of SnO\textsubscript{2} in all SnO\textsubscript{2}-containing clinoptilolite samples was additionally confirmed by the diffuse reflectance spectroscopy (DRS UV/Vis) analysis (Figure 3). A broad absorption peak centered at about 263 nm characteristic for pure SnO\textsubscript{2} (Figure 3a) is evident in all SnO\textsubscript{2}-containing clinoptilolite samples [29]. The translation, according to the Kubelka–Munk theory, revealed the bandgap of SnO\textsubscript{2} to be 3.38 eV, while the absorption intensities increased with the increase in Sn content. This can enhance the photocatalytic activity of the samples due to the enhanced formation of...
the electron-hole pairs by higher absorption of incoming irradiation. Similar observations have been reported for the incorporation of ZnO and CuO into clinoptilolite [13,30].

Once the presence of Sn-species was confirmed in the samples, the surface morphology of the SnO₂, CLI, and SnO₂-containing clinoptilolite was studied with scanning electron microscopy (SEM) to determine the location and size of tin oxide in the composites (Figure 4). Pure clinoptilolite shows typical crystal flake-like structures. Compared to pure CLI (Figure 4a), the SnO₂-clinoptilolite samples also showed the same crystal structures, which indicate the zeolite crystallites were not affected by grafting SnO₂. However, smaller particles that cover the clinoptilolite surface were seen and could be attributed to SnO₂ based on their size and spherical shape, which was distinguishingly different from that of clinoptilolite. The average size of SnO₂ agglomerates obtained from SEM images (histogram from 50 particles) was around 17 nm for pure SnO₂ and, similarly, 19 nm for SnO₂CLI25. Their distribution on the surface of clinoptilolite was seen as highly uniform. In addition, the further aggregation of the pure SnO₂ particles led to prolonged tube-like features (Figure 4b); these structures were absent in composites of higher SnO₂ loadings (SnO₂CLI25, Figure 4c), which showed a lesser aggregation of the SnO₂ particles in these samples.

Since tin oxide species were seen on the surface, we checked for the possible influence of these species on the textural properties with nitrogen physisorption, and the obtained results are given in Table 2. The Brunauer, Emmett, Teller (BET) surface area of CLI (24 m² g⁻¹) coincides well with the data reported elsewhere [31,32]. The porosity arises from interparticle voids among the different crystal structures present in the natural clinoptilolite (see XRD patterns above). It can be seen that the loading of CLI with SnO₂ increases specific surface area and total pore volume. The higher specific surface area provides better dispersion of Sn species, more active sites and should, hence, result in higher catalytic activity. The increase in BET surface area of the composites can be attributed to a partial dealumination of the clinoptilolite lattice during modification. Replacement of the metal...
cations present in the clinoptilolite channels by H\(^+\) resulted in more available space within the CLI lattice [28,33]. Moreover, an increase in the specific surface area with the increase in the Sn content was also evident. This could be attributed to the formation of an additional porous layer formed by amorphous SnO\(_2\) at the clinoptilolite surface.

| Sample          | \(S_{\text{BET}}\), m\(^2\) g\(^{-1}\) | \(S_{\text{ext}}\), m\(^2\) g\(^{-1}\) | \(V_1\), m\(^3\) g\(^{-1}\) | \(d\), Å |
|-----------------|--------------------------------------|--------------------------------------|-----------------|--------|
| CLI             | 24                                   | 22.129                               | 0.0988          | 162.58 |
| SnO2CLI05       | 36                                   | 32.421                               | 0.1266          | 146.87 |
| SnO2CLI15       | 62                                   | 61.179                               | 0.1562          | 134.48 |
| SnO2CLI25       | 78                                   | 78.545                               | 0.1712          | 105.83 |
| SnO\(_2\)       | 47                                   | 48.649                               | 0.1889          | 136.73 |

1 specific surface area based on Brunauer, Emmett, Teller (BET) theory determined in the P/P\(_0\) range corresponding to the increasing trend of Rouquerol plot; 2 external surface area \(S_{\text{BET}}-S_{\text{mic}}\); 3 total pore volume based on Barrett, Joyner and Halenda (BJH) adsorption analysis; 4 average pore size based on BJH adsorption analysis.

However, the average pore sizes in all samples were smaller than the particle aggregates that were seen in SEM imaging which suggests that SnO\(_2\) particles did not enter the pore system of clinoptilolite. The study of N\(_2\) adsorption/desorption isotherms (Figure 5) showed that, according to International Union of Pure and Applied Chemistry (IUPAC) classification, all presented isotherms could be classified as isotherms of type IV, ascribed to microporous adsorbents in which there exists a certain amount of impurities, especially clays, quartz, and amorphous glassy material. In addition, they showed a hysteresis loop of type H3, in the range \(p/p_0=0.5–1\). This feature, together with an upward deviation at high relative pressures, is typical for clinoptilolite structure and can be attributed to multilayer adsorption and capillary condensation either in mesopores of impurities (feldspars, quartz, etc.) or in the space between the zeolite crystallites [31].

![Figure 5](image.png)

**Figure 5.** Adsorption/desorption isotherms of the CLI, SnO\(_2\), and SnO\(_2\)-containing clinoptilolite samples.

Hence, to additionally confirm the state of Sn species and their location in the composites (either inside or outside the clinoptilolite’s pores) Sn K-edge Extended X-ray Absorption Fine Structure (EXAFS) analysis was used to determine the local structure around Sn cations in the SnO\(_2\)-containing clinoptilolite at different Sn loadings (SnO2CLI05, SnO2CLI15, and SnO2CLI25). Fourier transforms (FT) of the \(k^3\) weighted EXAFS spectra of the samples are shown in Figure 6. The spectrum of
crystalline SnO$_2$ with rutile structure is added for comparison. The Fourier transforms spectra reveal the contributions of consecutive shells of Sn neighbors up to about 4 Å. Qualitative comparison of the FT spectra indicates that the local structure around Sn cations in the samples was similar to that in the crystalline SnO$_2$. The lower signal in the interatomic distance (R) ranged from 2.5 Å to 4 Å which indicates that the number of neighbors in those coordination shells was smaller than in the bulk SnO$_2$, which indicates that Sn in the samples was present in the form of nanocrystalline SnO$_2$ particles with rutile structure and with different average nanoparticle sizes.

The quantitative analyses of Sn K-edge EXAFS spectra were performed to resolve structural parameters by comparing the measured signal with the model EXAFS spectrum, constructed with the FEFF6 program code [34], in which the photoelectron scattering paths were calculated ab initio from a tentative spatial distribution of neighbor atoms, based on the rutile type SnO$_2$ (cassiterite) crystal structure (space group P 42/m n m with unit cell parameters $a = b = 4.73727$ Å and $c = 3.18638$ Å, [35]). In this structure, Sn atoms are surrounded by six oxygen atoms in the nearest coordination shell and Sn and oxygen atoms in more distant coordination shells (two Sn at 3.19 Å and 8 Sn at 3.71 Å; 4 O at 3.59 Å and 4 O at 3.79 Å). The contributions of the first five coordination shells in the R range up to 4.0 Å were analyzed. The FEFF model comprised six single-scattering and all significant multiple-scattering paths from these neighbor shells. A distortion of the nearest oxygen octahedron (4+2) was allowed by introducing two Sn–O distances for the nearest coordination shell.

The crystalline SnO$_2$ spectrum was used to calibrate the model. A very good fit was obtained in the k range of 4.5–14.5 Å$^{-1}$ and R range of 1.2–4.0 Å (Figure 6), with just seven variable parameters: the amplitude reduction factor $S_0^2$ and shift of photoelectron energy origin $\Delta E_0$, lattice expansion ($\Delta r/r$), and the Debye–Waller factors of different scattering paths. The shell coordination numbers were
fixed at their crystallographic values. The quality of the fit is presented in Figure 7 and the best-fit parameters are listed in Table S3.

Figure 7. Attenuated total reflection (ATR)-FTIR spectra of (a) SnO₂ and (b) SnO₂-containing clinoptilolite. In (c) a closer range (1400–800 cm⁻¹) is shown.

The same EXAFS model was used to describe the Sn EXAFS spectra of the samples, except that the amplitude reduction factor S₀² was kept fixed at the value of crystalline SnO₂, while the number of Sn neighbors in the coordination shells beyond the nearest oxygen shell were allowed to vary. Octahedral coordination of the nearest oxygen neighbors was kept fixed.

In all three samples, Sn neighbors in the first five coordination shells of the SnO₂ cassiterite structure were found at the same distances (within error bars) and with the same or slightly larger Debye–Waller factors as in the crystalline SnO₂. However, in more distant coordination shells, the average Sn coordination numbers were significantly lower than in the bulk SnO₂ (about 50% of the value in bulk in the case of the SnO2CLI25 sample or even less for the second Sn neighbors at 3.71 Å in SnO2CLI15 and SnO2CLI05 samples). The results clearly indicate that SnO₂ nanoparticles with a cassiterite crystal structure were formed. The average size of nanoparticles was the smallest in the sample SnO2CLI05 with the smallest Sn loading.

We also tested eventual partial Sn incorporation into the zeolite structure. In the fitting of the second Sn coordination shell, the contributions of Si or Al neighbors backscattering were ruled out, judging from the unfavorable fit results (absence of Sn–O–Si and Sn–O–Al connections). We can conclude that Sn atoms were not incorporated into the zeolite crystal structure or directly bonded to the zeolite surface, or more precisely, that relative amount of such Sn cations was below the detection limit (~5%) of the EXAFS spectroscopy.

Since the nature and position of SnO₂ particles in the composites were determined, chemical and physical properties of the catalysts’ surface were studied; these can significantly affect the catalytic performance. First, attenuated total reflection (ATR-FTIR) spectroscopy was used to study the chemical nature of the surface. The results are shown in Figure 7. The presence of characteristics absorption bands of zeolite structure, such as the external tetrahedral linkage asymmetric stretching (at 1026 cm⁻¹) and linkages between SiO₄-tetrahedrons (at 792 and 608 cm⁻¹), confirmed the stability of zeolite structure during modification to SnO₂-loaded CLI. A lower intensity of band located at 1630 cm⁻¹, which belongs to vibrations of water molecules, can be attributed to lower hydration of the calcined samples with respect to the parent CLI. However, corresponding absorption bands of SnO₂ (~520 cm⁻¹, Figure 7a) were not noticeable probably due to the overlapping of bands and the strong intensity of the zeolite structure vibrations [15,36]. In addition, ATR-FTIR spectra indicate that the bands characteristic for asymmetric Si–O–Si stretching present at 1026 cm⁻¹ in CLI shifted to higher frequencies for HCLI (1052 cm⁻¹) and for additional 10 cm⁻¹, i.e., to about 1063 cm⁻¹ in SnO₂-containing samples. The first shift (for HCLI) can be ascribed to the dealumination process which strengthened the Si–O–Si bonds. According to Breck [37], the shift to even higher frequencies (i.e., 1063 cm⁻¹ in SnO₂-containing samples) can be ascribed to entering of transition metal cations into the zeolitic structure during the ion exchange process, yet leaving the zeolite structure unchanged. These observations confirm the partial loading of SnO₂ into the zeolite pore system. Such shifts have been observed in combination
with other nanoparticles supported on clinoptilolites, such as TiO$_2$, ZnO, SnO$_2$-ZnO, SnO$_2$-NiO, and CuO [14,17,30,38].

Lastly, the surface charge of the composites was studied since this property can largely influence the adsorption process in the composites. Zeta potential measurements (Figure 8) showed that CLI had a negative zeta potential in the whole pH range (pH = 1.5–11). For pure SnO$_2$, the point of zero charge (PZC) was nearly pH = 4.0, which is close to the literature data [39]. As expected, the surface charge of CLI was affected by loading with different amounts of SnO$_2$. PZC of SnO$_2$-clinoptilolite clinoptilolite was 2.66, 2.67, and 2.88 for SnO$_2$CLI05, SnO$_2$CLI15, and SnO$_2$CLI25, respectively. At pH value at which photocatalytic experiments were performed (pH = 5.6), surface SnO$_2$-containing clinoptilolite had negatively charged sites. Considering the positive charge of cationic MB dye, the adsorption was expected to be favored at this pH. Contrary, at pH below the PZC, the electrostatic repulsion between the positively charged surface of prepared photocatalysts and MB resulted in a decrease in their adsorption rate.

![Figure 8. Surface charges of CLI, SnO$_2$, and SnO$_2$-containing clinoptilolite samples.](image-url)

2.2. Photocatalytic Activity

The results of the photocatalytic tests are shown in Figure 9. It can be seen that the initial dark period (30 min) was enough to establish the adsorption equilibrium of SnO$_2$-containing clinoptilolite samples in contrast to raw CLI. It was, however, long enough to expose the discrepancy between the adsorption–desorption process and catalytic degradation process.
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(5.7), where the composite samples (and pure SnO 2) present more negative charge which more strongly attracts the positive MB dye.

Interestingly, the results show that the SnO 2 exhibited negligible adsorption capacity (4%) toward MB in the dark, while the degradation of MB under illumination with visible-light lamps was slightly higher (11%). The better adsorption capacity was obtained for raw CLI (10%), and the total degradation of MB in the presence of CLI was about 30%. By loading clinoptilolite with SnO 2, the photodegradation of the MB was increased, which demonstrated the importance of clinoptilolite support in adsorption of MB. It has also been found that the photodegradation of MB was affected by the content of SnO 2 onto clinoptilolite exhibiting the highest value for SnO2CLI25 (Figure 10). This can be explained by additional photocatalytic active sites on samples with a higher amount of SnO 2, as well as better dispersion of SnO 2 particles onto clinoptilolite with the higher specific surface area.

The reduction of methylene blue is a reversible and environment-dependent process, especially in oxygen-rich aqueous environments [40,41], which this system was. However, in semiconductor sensitized systems, the reduction of MB to its leuco form (LMB) is irreversible [42]. The test of removing the particles and continuing purging the solution with O 2 did not show a reverse of LMB to MB. Furthermore, we observed a slight blue shift of the 664 nm peak (monomer MB +) at longer irradiation times together with the rapid decrease in the dimer (MB +) 2 peak at 610 nm. These observations imply the destruction of the MB dye and its partial incorporation into the pores of the zeolite [43].

The importance of clinoptilolite support in the photodecolorization of MB was also reported by Nezamzadeh-Ejhieh et al. [13]. They reported that CuO incorporated onto micro-sized CLI showed better activity than raw CuO and CLI. This was explained by (1) the ability of zeolite to prevent aggregation of CuO particles by fixing them onto ion-exchange sites and (2) for the photocatalytic part, the ability of zeolite to distribute excited electrons in the conduction band of CuO into the network structure and thus prevent the recombination of electron-hole pairs. Similar observations were reported for photodecolorization of bromthymol blue by using supported ZnO onto CLI [14].

Because the composite SnO2CLI25 showed the highest final removal of MB dye, it was chosen as the catalyst for which additional characterization and optimization were conducted. 

Figure 9. Methylene blue (MB) photodegradation by using CLI, SnO2, and SnO2-containing clinoptilolite (a) and representation of the results with bars showing adsorption and degradation phase (b). Reaction conditions: \( m_{\text{cat}} = 10 \) mg, \( V = 50 \) cm\(^3\), \( O_2 \) flow = 25 cm\(^3\) min\(^{-1}\), \( T = 22 \) °C, \( C_0 = 10 \) mg dm\(^{-3}\).
were compared and are shown in Figure 11b. It can be seen that 10 mg was the optimal mass for the adsorption of MB was more a phenomena could be overcome with improved reactor design and the possible inclusion of UV-light photocatalyst. Hence, the effective surface area of the photocatalyst. Moreover, this can be explained by the aggregation of the photocatalyst particles which resulted in the reduction in the effective surface area of the photocatalyst. Additionally, as mentioned before, monomeric MB dye was partially incorporated in the zeolite’s pores, which can later be hard to remove. These phenomena could be overcome with improved reactor design and the possible inclusion of UV-light irradiation as the catalyst excitation source.

2.2.1. The Effect of the Amount of the Photocatalysts

The practical environmental application of photocatalyst requires the optimal amount of photocatalyst. Hence, the effect of the photocatalyst amount on the photodegradation of MB was studied. The results obtained by using different amounts (in the range of 7.5 to 40 mg) of the SnO2CLI25 sample are given in Figure 10. From the obtained results, it is clear that the final photodegradation of MB increased with the increase in the amount of the photocatalyst, reaching the highest value (75%) for the amount of 40 mg. However, the adsorption of MB was more affected by the amount of photocatalyst than the degradation of MB under illumination with visible-light lamps.

\[
A = (A_0 - A_\infty) \cdot \exp(-k^*t) + A_\infty
\]  

where \(A_0\) and \(A_\infty\) are initial and final (equilibrium) absorption values at 664 nm of the MB dye, respectively. The coefficients of correlation \((R^2)\) for such fits to data were all greater than 0.997. The reasoning for using such kinetics instead of the more commonly used \(A = A_0\exp(-k^*t)\) is in the fact that the MB evolution clearly shows a plateau is reached after prolonged illumination. This can be ascribed to the formation of additional intermediate products which are hard to degrade and additionally compete with pristine MB for adsorption/catalytic sites. Additionally, as mentioned before, monomeric MB dye was partially incorporated in the zeolite’s pores, which can later be hard to remove. These phenomena could be overcome with improved reactor design and the possible inclusion of UV-light irradiation as the catalyst excitation source.

The reaction rate constants derived from non-linear least-square fitting to the kinetic equation were compared and are shown in Figure 11b. It can be seen that 10 mg was the optimal mass for the fastest kinetics, and an additional increase in the catalyst mass led to a decrease in the photocatalytic speed. Naturally, this is a consequence of the aggregation of the photocatalyst particles which resulted in the reduction in the effective surface area of the photocatalyst. Moreover, this can be explained by the scattering of light and reduction in light penetration.
where M represents an active metal site to which the MB-dye preferentially adsorbs. However, the catalyst was still active in the degradation of MB under illumination with visible-light lamps. A decrease in catalytic activity in the photodecolorization of MB was also observed for CuO-supported nano-clinoptilolite [30].

2.2.2. Reusability of the Photocatalysts

The reusability of the photocatalyst for the MB degradation was also studied. The reused photocatalyst (SnO2CLI25) was washed with 0.01 M HNO3 three times (Figure 11a). Such catalysts were used again in the photodegradation test at optimized experimental conditions.

SEM images of reused photocatalyst (Figure 11c) show that morphology properties were preserved after the photodegradation test. In addition, EDX analysis of the reused photocatalyst showed that Sn content was almost unchanged (2.74 and 2.70 at.%, for fresh and reused SnO2CLI25, respectively). Besides, S (0.14 at.%) and Cl (0.12 at.%) were detected due to MB adsorption onto photocatalyst and, since the washing process did not remove the adsorbed dye completely (see discussion below), the detection of these elements was expected. Furthermore, ATR-FTIR analysis (Figure 11b) confirmed that the clinoptilolite structure was retained after the photocatalytic test, and also indicated the presence of derivatives of MB onto photocatalyst. The adsorbed MB dye was evident from the IR spectra by (i) the absorption of hydroxyl groups at 3441 cm\(^{-1}\) was broadening and offset. Since adsorption of MB dye is pH-dependent, increasing the pH would increase the negative potential of the zeolite surface as proven by zeta potential measurements and the decrease in the pH would result in the increased positive potential of zeolite. The pH-dependence of MB adsorption onto clinoptilolite can be described as:

\[
\text{M-OH} + \text{MB}^+ \rightarrow \text{M-O-MB} + \text{H}^+ \quad (2)
\]

\[
\text{M-O}^− + \text{MB}^+ \rightarrow \text{M-O-MB} \quad (3)
\]

where M represents an active metal site to which the MB-dye preferentially adsorbs. However, the zeolite-SnO\(_2\) composites show positive potential below pH = 3, hence for desorption, 0.01 M solution of HNO\(_3\) was chosen and proven effective (Figure 11a) for desorption of the dye from the clinoptilolite surface; MB can also be washed with ethanol, although in the materials in this study, this efficiency was very low, few percent per wash. This way, approximately 50% of the dye could be desorbed from the surface. The lack of incomplete desorption of the MB dye can be explained by the negative potential of clinoptilolite across the whole pH-range. Hence, even at pH = 2, the zeolite surface still attracts the positively charged MB molecules.

As can be seen from Figure 12, a decrease in the activity of catalyst during repeated cycles was observed. Adsorption of MB (initial dark period) was reduced, probably due to the blockage of the active sites on the clinoptilolite surfaces by the degradation products. N\(_2\)-isotherm pointed to a high number of available sites in the pristine zeolite (evident in the high nitrogen pressure range, where repulsive forces were predominant [44]), and these could be partially filled by the reaction intermediates, hindering multilayer adsorption which affected consecutive catalytic tests. However, the catalyst was still active in the degradation of MB under illumination with visible-light lamps. A decrease in catalytic activity in the photodecolorization of MB was also observed for CuO-supported nano-clinoptilolite [30].
Bahrami and Nezamzadeh-Ejhieh also reported a reduction in the photodecolorization of the BTB due to the adsorption of organic intermediates and by-products of the photodecolorization process in the cavities and on the catalyst (CuO- and ZnO-containing clinoptilolite) surface [14,30].

Figure 12. Reusability of the photocatalyst SnO2CLI25 in photodegradation of MB. The illumination part of the reaction, right of the dashed line, is fitted to Equation. 1 with a blue solid line. Reaction conditions: $V = 50$ cm$^3$, O$_2$ flow = 25 cm$^3$ min$^{-1}$, $T = 22$ °C, $C_0 = 10$ mg dm$^{-3}$, $m_{\text{cat}} = 10$ mg.

3. Materials and Methods

3.1. Preparation of Photocatalysts

In this study, natural zeolitic tuff (CLI) obtained from the Slanci deposit (Veliko selo, near Belgrade, Serbia) was used as a starting material. It consisted mainly of clinoptilolite (about 80 wt.%) and impurities including quartz (ca. 4 wt.%) and feldspar (ca. 16 wt.%). The particle size of the sample used was in the range of 63–125 µm. All chemicals used were used as purchased without further purification: SnCl$_2$·2H$_2$O (≥98%), NH$_4$OH (25%), NH$_4$CH$_3$COO (≥98%), C$_2$H$_5$OH (≥99.8%) was purchased from Sigma–Aldrich (St. Louis, MO, United States) and methylene blue (C$_{16}$H$_{18}$N$_3$SICl·xH$_2$O) from Merck (Kenilworth, NJ, United States). All aqueous solutions were prepared using Milli-Q deionized water.

3.1.1. Preparation of Pure SnO$_2$

The pure SnO$_2$ was prepared following the precipitation method described elsewhere [23,43]. In a typical synthesis procedure, 2.2820 g of SnCl$_2$·2H$_2$O was dissolved in 60 cm$^3$ of absolute ethanol, then NH$_4$OH solution (25%) was added dropwise with continuous vigorous stirring until the solution rich a pH = 8. The precipitated product was collected by filtration and suspended in a 4% CH$_3$COONH$_4$ solution for 30 min. Then, the precipitate was filtered, dried at 100 °C for 24 h and finally calcined at 400 °C for 2 h. The calcined powder (SnO$_2$) was then ground and used for the photocatalytic experiment.

3.1.2. Preparation of the SnO$_2$-Containing Clinoptilolite

Prior to its use CLI was washed several times with deionized water to remove water-soluble impurities and dried overnight at 105 °C. Then CLI was converted to hydrogen clinoptilolite (HCLI) by using a triple treatment of the CLI with 1 mol dm$^{-3}$ HCl at 100 °C for 4 h (solid/liquid ratio = 1:70) and treatment with 0.2 mol dm$^{-3}$ NH$_4$OH at 65 °C for 0.5 h (solid/liquid ratio = 1:30). Finally, the obtained product was separated and dried until constant weight. Further, HCLI was converted to SnO$_2$-containing clinoptilolite by using a slightly modified procedure (precipitation–deposition method followed by calcination step) given by Matatsushashi and Sowmiya [23,45]. HCLI was suspended in deionized water using solid/liquid ratio 1:100 and stirred for about 10 min. Then, the pH of the suspension was adjusted to 10 by dropwise addition of a 25% NH$_4$OH. The ethanol solution of SnCl$_2$
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containing different amounts of SnCl2·2H2O was dropwise under constant stirring to obtain 5, 15, and 25 wt.% Sn onto CLI. After repeated adjusting of pH to 10, and stirring for 24 h, the suspension was separated by centrifugation, washed with 4 wt.% CH3COONH4 solution and then with deionized water until chloride free. The obtained solids were dried overnight at 120 °C and then calcined at 400 °C for 2 h. Note, that although the measured contents of Sn differ from the nominal one (elemental analysis, Table 1) the prepared products were denoted as SnO2CLI05 (3 wt. % of Sn in CLI by el. analysis), SnO2CLI15 (10 wt. % of Sn in CLI by el. analysis) and SnO2CLI25 (15 wt. % of Sn in CLI by el. analysis).

3.2. Characterization of the Materials

The crystallinity of the prepared samples was studied by using an APD2000 Ital Structure diffractometer with CuKα radiation (λ = 0.15418 nm) in the 2θ range 5–60° with a step size of 2θ = 0.02°.

The thermal behavior of the samples obtained before the calcination step was examined by thermal analysis using an SDT Q600 simultaneous TGA–DTA instrument (TA Instruments). The samples were heated from room temperature to 800 °C at a heating rate of 10 °C min⁻¹ under synthetic air (100 mL min⁻¹).

To detect surface functional groups present in all prepared photocatalysts as well as spent photocatalysts, FTIR spectroscopy using attenuated total reflectance (ATR) apparatus were carried out by using an Alpha–Bruker FTIR spectrophotometer using a diamond crystal in a horizontal position. ATR-FTIR spectra were collected in the spectral region from 4000 to 400 cm⁻¹ with a resolution of 4 cm⁻¹.

The surface morphology of the samples was studied by SEM analysis using a Zeiss Supra 3 VP field emission gun (FEG) microscope, operating at 1 kV. The elemental composition of the samples was determined by the EDX analysis with INCA Energy System, attached to the above-described microscope, operating at 20 kV. Prior to the EDX analysis, samples were coated with carbon with a thickness of about 8–10 nm.

The textural properties of the samples were determined by the adsorption/desorption of nitrogen at –196 °C using an automatic sorption analyzer (ASAP 2020, Micrometrics, Norcross, GA, United States). Before the analysis, all samples were degassed overnight at 200 °C. The specific surface area of the sample (S_BET) was calculated according to the Brunauer, Emmett, Teller (BET) method from the linear part of the nitrogen adsorption isotherm [46]. The total pore volume (V_tot) was determined at P/P0 = 0.998.

The local atomic structure around Sn cations in the SnO2-containing clinoptilolite samples was analyzed by X-ray absorption spectroscopy. Sn K-edge EXAFS spectra of the samples were recorded at room temperature in the transmission detection mode at the F65 beamline of PETRA III, DESY, Hamburg, Germany.

The SnO2CLI5, SnO2CLI15, and SnO2CLI25 sample powders and crystalline SnO2 reference compounds were prepared as self-supporting homogenous pellets with total absorption thickness (µd) about 2 above the Sn K-edge. A Si (311) crystal monochromator was used with an energy resolution of about 3 eV at the Sn K-edge (29200 eV). Higher-order harmonics were removed by a flat Pt coated mirror placed in front of the monochromator. The intensity of the monochromatic X-ray beam was measured by three consecutive ionization detectors (5 cm long ionization chambers; the first filled with 400 mbar Kr and 600 mbar N2, the second and the third with 1000 mbar Kr). The samples were mounted on the sample holder placed between the first and second ionization detectors. The absorption spectra were measured in the energy region from −150 eV to +1000 eV relative to the Sn K-edge. In the edge region, equidistant energy steps of 0.25 eV were used, while in the EXAFS region, equidistant k steps of 0.03 Å⁻¹ were collected. The exact energy calibration was established with simultaneous absorption measurement on a 20-micron thick Sn metal foil placed between the second and the third ionization chamber. The absolute energy reproducibility of the measured spectra was ±0.03 eV.
analysis of EXAFS spectra was performed with the IFEFFIT program package [47] in combination with the FEFF6 program code [34] for ab initio calculation of photoelectron scattering paths.

The pH of the point of zero charge of the samples was determined as follows: 0.4 g of the sample (CLI, SnO₂, or SnO₂-containing clinoptilolite) was suspended in 100 cm³ of deionized water and sonicated for 2 h. Then, the pH of the suspension was adjusted accordingly by adding either 0.1 mol dm⁻³ HCl or 0.1 mol dm⁻³ NaOH solutions. Zeta potential measurements were performed by using the Zetasizer nano ZS instrument (Malvern, Malvern, United Kingdom). The pH value of each suspension was measured using a digital pH meter (Mettler Toledo, Columbus, OH, United States).

3.3. Photocatalytic Tests

Photocatalytic activity of the CLI, SnO₂, and SnO₂-containing clinoptilolite was evaluated in a batch reactor of volume 50 cm³. The amount of catalyst in the reaction was set to 10 mg and the concentration of methylene blue (MB) as a model dye was 10 mg dm⁻³, i.e., 0.5 mg of the dye was present in the reaction system. The reaction mixture was illuminated with visible-light lamps with an intensity of 8 mW cm⁻². The mixing of the system was provided by purging with O₂ at 25 cm³ min⁻¹ from the bottom of the cell (Figure S2).

First, 30 min of the dark period was monitored. Here the adsorption–desorption equilibrium was established. After that, the lights were turned on and the reaction was set to last for 3 h. In this part, the kinetics due to photocatalytic degradation of the dye could be evaluated and final removal assessed. The concentration of MB was measured spectrophotometrically, in the spectral range of 400–800 nm by UV Vis spectrophotometer (Agilent Technologies, Santa Clara, CA, United States) and measuring the absorbance at $\lambda_{\text{max}}$ of 664 nm for MB.

The effect of the photocatalyst amount was investigated by varying the amount of SnO₂CL115 (in the range of 7.5 to 40 mg). The photocatalytic test was performed using the procedure described in the previous section. For recycling tests, the used catalysts were first decanted and left to dry at room temperature. Then they were washed 3 times with 0.01 M HNO₃ to desorb the residual MB from the surface. They were finally dried at 90 °C for 1 h prior to the subsequent catalytic cycle.

4. Conclusions

Natural zeolite clinoptilolite was successfully modified with Sn using a precipitation–deposition method. The resulting SnO₂ nanoparticles were very small but formed aggregates of ~17 nm that were located on the surface of clinoptilolite without destroying the crystal structure of the zeolite. The composites were proven successful in degrading methylene blue dye even upon multiple recycling with acid treatment.

The increase in degradation efficiencies could be ascribed to the increased surface area with no destruction of the support due to the preferable location of SnO₂-aggregates on the surface of the zeolite. Increased adsorption and also increased degradation with higher Sn loadings were due to more negative surface potential, which was induced by the interaction between the zeolite and SnO₂.

These materials thus offer inexpensive and naturally abundant sources of catalysts able to operate under visible light illumination.

Supplementary Materials: The following are available online at http://www.mdpi.com/2073-4344/10/2/253/s1, Table S1. Chemical composition of the two clinoptilolites (wt.%) obtained by ICP-OES analysis, Table S2. Average elemental composition (wt.%) of CLI and SnO₂-containing clinoptilolite obtained by the EDX analysis, Figure S1. EDX spectra of the as-synthesized SnO₂ and SnO₂-containing composites, Table S3. Parameters of the nearest coordination shells around Sn atoms in Sn-modified CLI zeolites and reference SnO₂ compound with rutile crystal structure: N—the number of neighbor atoms, r—distance, $\sigma^2$—Debye–Waller factors, and R-factor—the goodness of fit parameter. The amplitude reduction factor and a shift of energy origin of photoelectron were fit in all the samples, the average obtained values were $S_0=0.96\pm 0.09$ eV and $\Delta E_0=2\pm 3$ eV. Uncertainty of the last digit is given in parentheses, Figure S2. A sketch of the photocatalytic set-up (a), a closer look at the internal part of the reactor (b) and a photograph of the whole set-up (c). On the right, a 3D model of the reactor with one side open.
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