Investigation of Ferrochromium Wastes for their Hazardous Hexavalent Chromium Content

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(First received 15 June 2021 and in final form 21 August 2021)

DOI: 10.31590/ejosat.952386

ATIF/REFERENCE: Acar, İ. (2021). Investigation of Ferrochromium Wastes for their Hazardous Hexavalent Chromium Content. European Journal of Science and Technology, (27), 204-209.

Abstract

In this study, ferrochromium slag (FS) and electric-arc furnace dust (EAFD) samples were examined for their hazardous hexavalent chromium content mainly by X-ray photoelectron spectroscopy (XPS) analysis. According to the experimental results, magnesium-silicate constitutes the main framework of both FS and EAFD samples. Highly crystalline FS contains 3.11%Cr$_2$O$_3$ as spinel phases while this constituent is much higher, 15.44%, in EAFD. The general XPS spectra show that both samples mainly consist of O, Mg and Si elements. In addition, FS also includes aluminum while carbon, calcium and potassium are the other major elements for EAFD. The peaks appeared in the range of 577-589 eV confirm chromium content of the samples. Besides, the peaks appeared at 1022 and 1045 eV point out zinc content of EAFD. Based on the XPS atomic ratio, Cr contents of FS and EAFD were obtained as 0.505 and 0.987%, respectively while EAFD further contains 2.594% Zn. The detailed XPS analysis indicates that chromium exists as two oxidation states, trivalent and hexavalent forms at 2p orbitals. Specifically, for 2p$_{3/2}$ orbitals, trivalent and hexavalent chromium were assigned at 577.1 and 579.2 eV, respectively whereas 586.9 and 588.9 eV for 2p$_{1/2}$ orbitals. Furthermore, the corresponding intensities of these peaks are much higher for EAFD. By comparing the peak intensities, the relative proportions of Cr (VI) were calculated as 49.44 and 49.53% for FS and EAFD, respectively. The leaching tests show that solubilities of all the metals examined except Cr are far below the toxicity limits for hazardous waste and even below the ranges of non-hazardous waste. Specifically, chromium contents of the filtrates were determined as 0.862 and 268.050 mg/l for FS and EAFD, respectively. Considering the toxicity limits of chromium, EAFD can be surely labeled as a highly hazardous waste whereas, on the other hand, FS is within the limits of non-hazardous waste. In conclusion, it can be said that the hazardous hexavalent chromium content must be reduced to immobile trivalent forms for the dust sample prior to be disposed in landfills.

Keywords: Hexavalent chromium, Ferrochromium slag, Electric-arc furnace dust, Hazardous waste, XPS.

Ferrokrom Atıklarının Tehlikeli Altı Değerlikli Krom İçeriği Açısından İncelenmesi

Öz

Bu çalışmada, ferrokom cürufu (FS) ve elektrik-ark firmı tozu (EAFD) numuneleri esas olarak X-ışını fotoelektron spektroskopisi (XPS) analizi ile tehlikeli altı değerlikli krom içeriğini açısından incelenmiştir. DeneySEL sonuçlara göre, magnezyum-silikat, hem FS hem de EAFD numunelerinin ana çerçevesini oluşturmaktadır. Yüksek kristal FS, spinel fazlar olarak %3,11 Cr$_2$O$_3$ içerikli, bu bileşen EAFD içerisinde çok daha yüksekliktir (%15,44). Genel XPS spektrumları, her iki numunenin de esas olarak O, Mg ve Si elementlerinden olduğu görülmüştür. Ek olarak, FS ayrıca alüminyum içerikli, karbon, kalsiyum ve potasyum EAFD için diğer ana elementlerden 577-589 eV aralığında görünen pikler, numunelerin krom içeriğini doğrulamaktadır. Ayrıca, 1022 ve 1045 eV'de ortaya çıkan pikler, EAFD'nin çok değerlikli içeriğine işaret etmektedir. XPS atom oranına göre, FS ve EAFD'nin Cr içerikleri sırasıyla %0,505 ve %0,987 olarak elde edilmiştir, EAFD ayrıca %2,594 Zn içerir. Ayrıca XPS analizi, kromun 2p orbitallarında üç değerlikli ve altı değerlikli formlar olmak üzere iki oksidasyon durumunda bulunduğunu göstermektedir. Spesifik olarak, 2p$_{3/2}$ orbitalları için üç değerlikli ve altı değerlikli krom sırasıyla 577,1 ve 579,2 eVde, 2p$_{1/2}$ orbitalları için ise 586,9 ve 588,9 eV'de belirlenmiştir. Ayrıca, bu piklerin karşılık gelen şiddetleri EAFD için çok daha yüksekliktir. Pik şiddetleri karşılaştırıldığında, Cr (VI) nişpi oranları FS ve EAFD için sırasıyla %49,44 ve %49,53 olarak hesaplanmıştır. Liç testleri, Cr hariç incelenen tüm metallerin çözünürlüklerinin, tehlikeli atık toksisite sınırlarının çok altında ve hatta tehlikeli olmayan atık aralıklarının dahi altında olduğunu göstermektedir. Spesifik olarak, filtratların krom içeriği FS ve EAFD için sırasıyla 0,862 ve 268,050 mg/l olarak belirlenmiştir. Kromun toksisite limitleri göz önüne alındığında, EAFD kesinlikle çok tehlikeli bir atık olarak etiketlenebilirken, FS ise tehlikesiz atık sınırları içindedir. Sonuç olarak,
European Journal of Science and Technology

1. Introduction

Ferrochromium consisting mainly of chromium, iron and carbon is one of the most widely used alloy for the manufacturing of stainless steel. The two by-products resulted from the ferrochrome production are ferrochromium slag (FS) and electric-arc furnace dust (EAFD) (Aigbe and Osibote, 2020; Bulut et al., 2009; Erdem et al., 2005; Jena and Panigrahi, 2021).

The main bulk waste, FS is generated from the successive carbothermal process in electric-arc furnaces and mostly forms as a dense rock with high degree of crystallinity due to the slow cooling in the air (Gencel et al., 2013; Sekhar Biswal et al., 2021).

Depending on the mineralogy and chemistry of feed materials, approximately 1.1-1.6 tons FS is resulted for a ton of ferrochrome produced. Accordingly, the annual production of FS is estimated around 12.5 million tons with the increasing rate of 2.8-3.0% by weight in every year (Nath, 2018). A small percentage of this finds applications in construction industry as mostly aggregate but the vast majority is held in dumps (Bulut et al., 2009; Gencel et al., 2013; Nath, 2018). However, this can be regarded as environmentally undesirable approach due to its high levels toxic elements (Bulut et al., 2009; Sohel et al., 2021; Panda et al., 2013).

EAFD, on the other hand, is generated by the subsequent volatilization, oxidation and solidification of metals such as Zn, Pb, Cd, Ni, and Cr in the form of a dust (Arnold et al., 2017). In contrast to FS, only 10-20 kg EAFD is resulted per a ton of ferrochrome produced. Approximately 30% of EAFD finds application as source materials for the recovery of metals, while the remaining 70% is disposed in landfills (Arnold et al., 2017; Bulut et al., 2009; Cholake et al., 2018). EAFD also contains high levels of extractable toxic elements such as Cr, Zn, Pb, Ni and Cd, and has been categorized as hazardous waste due to its chemical, physical and leachability properties. Accordingly, its disposing in landfills without any pre-treatments such as detoxification and immobilization processes is strongly prohibited by a wide range of countries (Bulut et al., 2009; Fernández Pereira et al., 2009; Shahid et al., 2017).

One of the most hazardous elements present in both FS and EAFD is chromium, particularly its hexavalent state. Despite its several oxidation states, the two most common and stable forms are hexavalent and trivalent chromium. Unlike trivalent chromite, hexavalent chromate has a potential health hazard mainly due to its high mobility, extreme toxicity and carcinogenic property. Therefore, the main detoxification mechanism is based on the reduction of hazardous hexavalent chromate to its immobile trivalent chromite form (Bulut et al., 2009; Chagas et al., 2019; Dhal et al., 2013; Fares et al., 2021; Fernández Pereira et al., 2009; Panda et al., 2013; Shahid et al., 2017).

In this study, a slag and an electric-arc furnace dust resulted from a ferrochrome production process were examined for their hexavalent chromium content mainly by X-ray photoelectron spectroscopy (XPS) analysis.

2. Material and Method

2.1. Materials

The ferrochrome slag (FS) and electric-arc furnace dust (EAFD) samples used were resulted from the ferrochrome production plant of ETI KROM INC. located in Kovancilar, Elazığ, Turkey. In this plant, after the removal of ferrochromium metal from the slag in liquid state, the molten slag is slowly cooled in air. This results in a highly crystalline irregularly-shaped lumpy material with approximately 4-5 cm in size (Erdoğan, 2011; Gencel et al., 2013; Nath, 2018). Therefore, prior to the instrumental analyses, FS sample was crushed and ground to particle size finer than 100 μm.

2.2. Methods

2.2.1. Characterization of the Materials

The main oxide composition was determined by a Spectro IQ X-ray fluorescence analyzer. The particle size distribution of EAFD was measured with wet-sieving analysis using Tyler sieve series from 425 μm down to 38 μm. The major crystalline inclusions of the samples were examined by a PANalytical Empyrean X-ray diffractometer (XRD) operating with Cu-Kα radiation (λ=1.54060 Å) at 40 kV and 40 mA over 10-70°. Thermogravimetric analyses were carried out with a TGA 1500 instrument in nitrogen atmosphere.

2.2.2. XPS Analysis

XPS analysis can be considered as an accurate method used to determine oxide state of transition metals (Greunz et al., 2018). Therefore, in this study, XPS analysis was conducted by a Specs-Flex instrument for the determination of hexavalent chromium contents of FS and EAFD samples. General survey spectra were recorded for the binding energy of 0-1387 eV at 100 eV pass energy while narrow range spectra for chromium element were acquired in the range of 564-604 eV at the pass energy of 40 eV.

2.2.3. Leaching Test

Leachability characteristics of FS and EAFD samples were examined for eight elements in accordance with TS EN 12457-4 leaching test (Acar and Atalay, 2013). Heavy metal contents of the filtrates were determined by an Agilent 7800 inductively coupled plasma-mass spectroscopy. The regulation of hazardous waste in Turkish standard (Appendix 11-A) was used to evaluate the obtained heavy metal contents for their toxicity limits (Acar and Atalay, 2013).

3. Results and Discussion

3.1. The Results of Material Characterization

The major oxide compositions of the samples are given in Table 1.

Table 1 The main chemical compositions of FS and EAFD

| Constituents (%) | FS    | EAFD  |
|-----------------|-------|-------|
| MgO             | 38.31 | 15.13 |
| SiO₂            | 32.56 | 18.88 |
| Al₂O₃           | 23.01 | 3.65  |
| Cr₂O₃           | 3.11  | 15.44 |
| Fe₂O₃           | 1.29  | 7.37  |
| CaO             | 1.22  | 6.26  |
According to Table 1, MgO, SiO$_2$ and Al$_2$O$_3$ constitute the main chemical composition of FS which also includes 3.11% Cr$_2$O$_3$ and very low amounts of Fe$_2$O$_3$ and CaO contents. Similar chemical compositions with varying Cr$_2$O$_3$ contents were also reported in the literature (Erdoğan, 2011; Gencel et al., 2013; Jena and Panigrahi, 2021). On the other hand, EAFD contains much higher Cr$_2$O$_3$ content, 15.44%, and lower SiO$_2$, MgO and Al$_2$O$_3$ contents with respect to FS, all of which are correlated with the literature investigations (Bulut et al., 2009; Kumar B et al., 2020; Mishra et al., 2020).

The particle size distribution of EAFD sample is given in Figure 1. As seen in Figure 1, almost all of the particles pass through 425 µm while 80% ($P_{80}$) and 50% ($P_{50}$) cumulative passing sizes by weight were determined to be approximately 143 µm and 79 µm, respectively. Similar particle size distribution was also reported in the literature (Bulut et al., 2009).

![Figure 1 The particle size distribution of EAFD](image)

The XRD pattern of FS and EAFD samples are illustrated in Figure 2. As seen in Figure 2-(a), FS sample mainly consists of magnesium-silicate and magnesium-aluminate minerals. In addition, it exhibits high crystallinity probably due to the slow air cooling. Specifically, forsterite ((Mg,Fe)$_2$SiO$_4$), Fe-Mg-Cr-Al-spinels (MgAl$_2$O$_4$) and enstatite (MgSiO$_3$) phases were determined from the XRD analysis. The results obtained are well correlated with the literature studies (Acar, 2020; Erdoğan, 2011; Gencel et al., 2013; Nath, 2018).

On the other hand, Figure 2-(b) indicates that the main crystalline inclusions of EAFD were determined as forsterite (Mg$_2$SiO$_4$) and cristobalite low (SiO$_2$), and also includes magnesium dichromium oxide (Cr$_2$MgO$_4$), calcium silicide (CaSi), magnesium silicide (Mg$_2$Si), fayalite (Fe$_2$SiO$_4$) and stishovite (SiO$_2$). Parallel crystalline phases were also reported in the literature (Mishra et al., 2020).

According to the thermogravimetric analyses conducted, FS exhibited almost no weight reduction, only 0.065% up to the highest temperature tested (Gencel et al., 2013). On the other hand, EAFD showed 8.22% weight reduction probably due to its unburn coke inclusion. High loss on ignition of electric-arc furnace dust was also reported in the literature (Fernández Pereira et al., 2009).

![Figure 2 Mineralogical compositions of FS (a) and EAFD (b)](image)

3.2. The Results of XPS Measurements

Figure 3 displays the general XPS spectra of FS and EAFD samples. As seen in Figure 3-(a), the peaks of O, Mg, Si, Al and Ca elements were mainly determined for FS. In addition, low intensity peaks appeared at 575 and 588 eV confirm chromium inclusion of FS sample.

Similarly, the peaks of O, Mg, Si and Ca elements except Al can be clearly seen for EAFD in Figure 3-(b). Different to FS, the general XPS spectra of EAFD also includes the peaks of K, Na, Cl and S elements probably resulted from burning of coke. Moreover, higher intensity peaks at 577 and 588 eV with respect to FS confirm the higher chromium content of EAFD sample. In addition to chromium, another heavy metal element, zinc was also determined from the peaks appeared at 1022 and 1045 eV.

XPS atomic ratios of FS and EAFD samples are given in Tables 2 and 3, respectively. According to Table 2, consistent with the XRF results, O, Mg, Si and Al constitute the main elemental composition of FS sample. Inconsistent Ca and C contents with regard to the XRF and TGA results can be attributed to the peak overlaps among Ca, C and Mg.

Table 3 indicates that the major elements detected for EAFD sample are O, Mg, Si, C, Ca and K. In addition, EAFD also has a high Zn content, 2.594%, and about two times higher Cr content, 0.987%, with respect to FS sample.
attributed to the binding energy ranges of 586.66-586.98 and 588.77-589.68 eV (Greunz et al., 2018; Murphy et al., 2009).

Table 3 XPS atomic ratio of EAFD

| Elements | Orbitals | Positions (eV) | Atomic ratio (%) |
|----------|----------|---------------|------------------|
| O        | 1s       | 531           | 49.953           |
| Mg       | 2s       | 88            | 17.115           |
| Si       | 2p       | 101           | 7.236            |
| S        | 2p       | 169           | 1.624            |
| Cl       | 2p       | 199           | 1.015            |
| Ca       | 2p       | 351           | 5.965            |
| Na       | 1s       | 1072          | 1.369            |
| K        | 2p       | 293           | 4.276            |
| Cr       | 2p 3/2   | 577           | 0.987            |
| Zn       | 2p 3/2   | 1022          | 2.594            |
| C        | 1s       | 285           | 7.865            |

Table 2 XPS atomic ratio of FS

| Elements | Orbitals | Positions (eV) | Atomic ratio (%) |
|----------|----------|---------------|------------------|
| O        | 1s       | 530           | 56.200           |
| Mg       | 2s       | 87            | 10.888           |
| Si       | 2p       | 100           | 9.212            |
| Al       | 2p       | 72            | 5.595            |
| Ca       | 2p       | 350           | 7.321            |
| Cr       | 2p 3/2   | 575           | 0.505            |
| C        | 1s       | 283           | 10.142           |

Figure 4 illustrates the detailed narrow range spectra of FS and EAFD for chromium. As seen in Figure 4, two oxidation states, trivalent and hexavalent chromium were determined at 2p_{3/2} and 2p_{1/2} orbitals. According to the literature investigations, for Cr 2p_{3/2} orbitals, trivalent chromium forms of Cr₂O₃ and CrCl₃ are characterized with the peaks appeared at 576.2-576.5 and 577.2 eV, respectively. On the other hand, hexavalent chromium forms of CrO₃ and K₂Cr₂O₇ correspond to higher binding energies, 578.1 and 579.2 eV, respectively. For Cr 2p_{1/2} orbitals, trivalent and hexavalent chromium forms are sequentially

As seen in Figure 4-(a), for FS, consistent with the literature studies, trivalent chromium was assigned at 577.1 eV with the peak intensity of 5.257×10⁴ cps while hexavalent form corresponded to 579.2 eV with 5.109×10⁴ cps for 2p_{3/2} orbitals. On the other hand, lower intensity peaks were obtained for 2p_{1/2}
orbitals. Specifically, trivalent and hexavalent oxidation states were sequentially determined at 586.9 and 588.9 eV with the respective intensities of 5.032x10⁴ and 4.954x10⁴ cps. By comparing the intensities of the peaks examined, the relative proportions of Cr (III) and Cr (VI) for FS were calculated as 50.56 and 49.44%, respectively.

According to Figure 4-(b), almost the same binding energies with respect to FS were determined for EAFD sample. However, in this case, much higher peak intensities were observed for the corresponding orbitals and oxidation states. Specifically, for 2p³/₂ orbitals, trivalent and hexavalent forms were assigned with 8.201x10⁴ and 8.067x10⁴ cps, respectively. For 2p₁/₂ orbitals, the intensity values of 8.182x10⁴ and 8.011x10⁴ cps were obtained for the respective trivalent and hexavalent chromium. Accordingly, the relative proportions of Cr (III) and Cr (VI) for EAFD were calculated as 50.47 and 49.53%, respectively. In other words, almost half of the total chromium content for both FS and EAFD are present as hazardous hexavalent forms.

3.3. The Results of Leaching Tests

The leachability characteristics of heavy metals from FS and EAFD are given in Table 4.

| Element | Solubility (mg/l) | Toxicity limits (mg/l) (Appendix 11-A) (Acar and Atalay, 2013) |
|---------|------------------|---------------------------------------------------------------|
| FS      | EAFD             | IWᵃ                | NHWᵇ                | HWᶜ                |
| Cd      | 0.00003          | 0.00023            | ≤ 0.004             | 0.004-0.1         | < 0.1-0.5          |
| Co      | 0.00072          | 0.00005            | Niᵃ | Niᵈ | Niᵈ |
| Cr      | 0.862            | 268.050            | ≤ 0.05             | 0.05-1            | < 1-7              |
| Cu      | 0.0025           | 0.0037             | ≤ 0.2              | 0.2-5             | < 5-10             |
| Ni      | 0.0154           | 0.00076            | ≤ 0.04             | 0.04-1            | < 1-4              |
| Pb      | 0.00090          | 0.0056             | ≤ 0.05             | 0.05-1            | < 1-5              |
| Zn      | 0.0191           | 0.162              | ≤ 0.4              | 0.4-5             | < 5-20             |
| Mn      | 0.028            | 0.0021             | Niᵃ | Niᵈ | Niᵈ |

ᵃInert waste.
ᵇNon-hazardous waste.
ᵈHazardous waste.
⁴Not included.

According to Table 4, both FS and EAFD are far from being considered as hazardous waste in terms of the leaching capacities of all the metals examined except Cr. In addition, their solubilities are even below the toxicity limits for non-hazardous waste. In particular, solubilities of chromium were determined as 0.862 and 268.050 mg/l for FS and EAFD, respectively. In other words, chromium solubility of EAFD is more than 300 times higher than that of FS based on the TS EN 12457-4 leaching test. This result cannot be explained by only the chromium contents of FS and EAFD, which are 3.11 and 15.44%, respectively. In fact, it can be attributed to the spinel phase of FS where most of the chromium is immobilized. This results in a slight chromium release from FS (Panda et al., 2013). On the other hand, chromium is easily leached from the weakly crystalline magnesium dichromate oxide phase of EAFD. Therefore, EAFD can be regarded as a highly hazardous waste while FS is within the limit of non-hazardous waste according to the toxicity limits (Bulut et al., 2009; Fernández Pereira et al., 2009; Panda et al., 2013).

4. Conclusions and Recommendations

Overall results have shown that although the ferrochromium slag and the electric-arc furnace dust examined have similar Cr(III)/Cr(VI) ratio, mainly weak crystallinity makes the dust sample a highly hazardous waste while the slag sample can be considered as a non-hazardous waste due to its strong chromium bearing spinel phase. Therefore, hazardous hexavalent chromium must be reduced to much less dangerous immobile trivalent forms for the dust sample prior to disposed in landfills.

5. Acknowledge

The technical supports by East Anatolia High Technology Application and Research Center (DAYTAM) and ETI KROM INC. are gratefully acknowledged.

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