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Ambient Air Pollution and Hospital Admissions for Peptic Ulcers in Taipei: A Time-Stratified Case-Crossover Study

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Abstract: Very few studies have been performed to determine whether there is a relationship between air pollution and increases in hospitalizations for peptic ulcer, and for those that have occurred, their results may not be completely relevant to Taiwan, where the mixture of ambient air pollutants differ. We performed a time-stratified case-crossover study to investigate the possible association between air pollutant levels and hospital admissions for peptic ulcer in Taipei, Taiwan. To do this, we collected air pollution data from Taiwan’s Environmental Protection Agency and hospital admissions for peptic ulcer data for the years 2009–2013 from Taiwan’s National Health Insurance’s research database. We used conditional logistic regression to analyze the possible association between the two, taking temperature and relative humidity into account. Risk was expressed as odds ratios and significance was expressed with 95% confidence intervals. In our single pollutant model, peptic ulcer admissions were significantly associated with all pollutants (PM$_{10}$, PM$_{2.5}$, SO$_2$, NO$_2$, CO, and O$_3$) on warm days (>23 °C). On cool days (<23 °C), peptic ulcer admissions were significantly associated with PM$_{10}$, NO$_2$, and O$_3$. In our two-pollutant models, peptic ulcer admissions were significantly associated NO$_2$ and O$_3$ when combined with each of the other pollutants on warm days, and with PM$_{10}$, NO$_2$, and O$_3$ on cool days. It was concluded that the likelihood of peptic ulcer hospitalizations in Taipei rose significantly with increases in air pollutants during the study period.

Keywords: ozone; nitrogen dioxide; air pollution; peptic ulcer, case-crossover; hospital admissions

1. Introduction

The association between both short- and long-term exposure to air pollutants and daily mortality and hospitalizations attributed to diseases of the cardiovascular system and diseases of the lungs is well known [1–5]. Sustained long-term exposure to air pollution has been found to substantially lower life expectancy [6–8].

Several years ago, no correlation had been found between air pollution and hospitalization for diseases of the gastrointestinal track [9,10]. The relationship was once considered so unlikely that some studies investigating the effect of air pollution used intestinal diseases as a negative control [11,12]. However, there has been a recent increased interest in the effects of atmospheric pollutants on digestive
diseases, including enteritis [13], appendicitis [14–16], inflammatory bowel diseases [17,18], non-specific abdominal pain [19], and peptic ulcer disease [20–22].

A peptic ulcer is a disorder of the gastroduodenal mucosa in which the epithelial cells rupture, causing pain and bleeding, and sometimes perforation [22]. Use of non-steroidal anti-inflammatory drugs (NSAIDs) and low-dose aspirin as well as infection by *Helicobacter pylori* have been clearly associated with increased risk of this disease [23,24]. However, a significant proportion of peptic ulcer cases cannot be explained by these well-known risk factors [25,26]. Interestingly, gastrointestinal bleeding has been found to occur more frequently in winter months, a finding that suggests its incidence could be affected by environmental factors [27].

Exposure to air pollutants have been shown to alter intestinal immunity [28], increase intestinal permeability [29,30], and alter gut microbial composition [28,31], all factors that could contribute to the development of peptic ulcers. Three recent epidemiologic studies have directly examined the specific relationship between air pollution and hospital admissions for peptic ulcer [20–22]. The results of these studies may not, however, be clearly applied to Taiwan, a region where air pollutant mixtures may differ. Therefore, by collecting official air pollutant data and national health insurance claim data, we performed a case-crossover study to investigate the possible association between short-term levels of various air pollutants and daily hospital admissions for peptic ulcers in Taipei, Taiwan.

2. Materials and Methods

2.1. Taipei City

Taipei City, with a population of 2.64 million, is the largest urban area in Taiwan. Its climate is subtropical with an average temperature of 23 °C. Although its extensive subway was created to reduce traffic congestion and air pollution, vehicle exhaust, especially from automobiles and motorcycles, remain the city’s primary source of air pollution [32].

2.2. Hospital Admission Data

For this study, we collected daily counts of hospital admissions (including emergency room admissions and scheduled admissions) with a primary diagnosis of peptic ulcer (ICD-9 531-532) from Taiwan’s 2009 to 2013 National Health Insurance (NHI)’s Research Database (NHIRD), which includes insurance claims for almost all people in Taiwan. Taiwan’s National Health Insurance program, which was implemented on March 1, 1995, provides compulsory universal health insurance to all of Taiwan’s residents, covering about 98% of the population who receive care from 93% of the medical institutions and private facilities on the island. The NHI hospital dataset includes data elements such as birth day, gender, the date of admissions and discharge, and the International Classification of Diseases, 9th revision, Clinical Modification (ICD-9 CM) code for each admission. This dataset has been used as the basis for many past epidemiological studies and has been proved to be accurate and reliable [33]. This dataset, as well as the quality of its data, has been described previously [34]. This study was reviewed and approved by the Institutional Review Board at Kaohsiung Medical University Hospital (KMUH-IRB-exempt-20130018).

2.3. Air Pollution and Meteorological Data

As described previously [16], we collected air pollution data obtained from six fixed air quality monitoring stations located in Taipei from Taiwan’s Environmental Protection Administration. These monitoring stations measured hourly data of six air pollutants, including carbon monoxide (CO) (by nondispersive infrared photometry), sulfur dioxide (SO₂) (by ultraviolet fluorescence), particulate matter (PM₁₀ by beta-ray absorption and PM₂.₅ by tapered element oscillating microbalance method), nitrogen dioxide (NO₂) (by ultraviolet fluorescence), and ozone (O₃) (by ultraviolet photometry). The completeness of hourly data in these monitoring stations was above 97% for all pollutants during the study period (the mean was 97.7%). Hourly air pollution data from each of the monitors were
collected for each day to calculate the mean daily levels. The daily exposure to each air pollutant was estimated by averaging hourly data across the stations, and from these values we calculated the daily average concentrations. We used the mean concentrations from the monitoring stations to represent the daily levels of Taipei City. Meteorological data including daily mean temperature and humidity were provided by the Central Weather Bureau’s Taipei observatory.

2.4. Statistical Analysis

This study used a case-crossover design [35–37], a good alternative to Poisson time-series regression models for evaluating the health effects of short-term increases in various pollutants, including those found in ambient air [35,38].

The study divided the observation period into several months. In this time-stratified approach [38], days falling on the same day of the week as the index day were defined as referent days in each month. We compared the case-period levels of various air pollutants with levels of those pollutants on all referent days. This approach is thought to reduce bias due to the non-stationarity of air pollution data [39,40]. In addition, the case-crossover design and the time-series approach yielded almost identical results [41,42]. Exposure to air pollutants was measured using the 2-day lagged cumulative moving average of air pollutant levels because previous studies have associated increased hospital admissions with high air pollutant levels on the same day of admission or over the previous two days [43,44]. Taipei City has a subtropical climate with no clear four-season cycle [32]. Hence, in this study, the possible interaction of seasonality on the effects of air pollutants was not considered; however, temperature was used instead. We analyzed our data by warm days (days with a mean temperature above 23 °C) and cool days (days with a mean temperature below 23 °C) only.

We performed conditional logistic regression to analyze the relationship between air pollutant levels and hospitalizations. Weights were equal to admissions counts on a given day. Risk was expressed as odds ratio (ORs) and significance was expressed as 95% confidence intervals (CIs). Exposure levels were entered as continuous variables. Meteorologic variables, such as temperature and humidity, which might play a confounding role, were considered in analyses. We adjusted for same-day average temperatures and humidity in the model. In all analyses, we modeled the same-day average temperature (lag 0) as a quadratic function, same-day average humidity (lag 0), and pollutants (the 3-day moving average from current day to previous 2 days concentration) as a linear function. ORs for the various air pollutants were calculated for increases in the interquartile range (IQR; between the 25th and the 75th percentiles). All statistical analyses were performed with the aid of SAS (version 9.4; SAS Institute, Inc., Cary, NC, USA).

3. Results

In total 23,205 patients were admitted for peptic ulcers in Taipei during the 5-year study period. As can be seen in Table 1, a summary of admissions characteristics and corresponding environmental data, peptic ulcer admissions averaged 12.71 per day over the entire study period, among which 62.54% were gastric ulcer and 37.46% were duodenal ulcer; 62.79% of the admissions were male and 37.21% were female. The average age was 65.77 ± 17.71 years. The mean daily count of admissions were 0.09, 5.37, 2.44, and 4.83 for ages <15, 15–64, 65–74, and >74 years, respectively.
**Table 1.** Distribution of daily peptic ulcer admissions, weather, and air pollutant concentrations in Taipei, Taiwan, 2009–2013.

| Variable               | Median a | Mean       | Maximum |
|------------------------|----------|------------|---------|
| PM$_{10}$ (µg/m$^3$)   | 42.51 (31.21–57.28) | 47.09 | 205.35 |
| PM$_{2.5}$ (µg/m$^3$)  | 25.24 (18.59–34.69)  | 27.80 | 140.54 |
| SO$_2$ (ppb)           | 2.80 (2.11–3.82)      | 3.08  | 9.04   |
| NO$_2$ (ppb)           | 22.01 (18.22–26.40)   | 22.73 | 61.94  |
| CO (ppm)               | 0.56 (0.44–0.71)      | 0.61  | 1.99   |
| O$_3$ (ppb)            | 23.29 (17.74–30.57)   | 24.69 | 63.15  |
| Temperature (°C)       | 23.91 (19.08–28.60)   | 23.48 | 33.18  |
| Humidity (%)           | 73.17 (66.64–80.35)   | 73.06 | 93.78  |
| Daily hospital admissions Peptic ulcer (n = 23,205) | 12 (10–16) | 12.71 | 32 |
| Gastric ulcer (n = 14,513) | 8 (5–10) | 7.96 | 20 |
| Duodenal ulcer (n = 8692) | 5 (3–6) | 4.77 | 15 |
| Male (n = 14,571)      | 8 (6–10) | 7.99 | 21 |
| Female (n = 8634)      | 4 (3–6)  | 4.73 | 17 |
| Age <15 (n = 163)      | 0 (0–0)  | 0.09 | 3 |
| Age 15–64 (n = 9787)   | 5 (3–7)  | 5.37 | 17 |
| Age 65–74 (n = 4453)   | 2 (1–3)  | 2.44 | 11 |
| Age >74 (n = 8802)     | 5 (3–6)  | 4.83 | 15 |

a median (25–75 percentile).

Pearson correlation coefficients among the air pollutants are presented in Table 2. A single pollutant model was used to analyze the associations between various ambient air pollutant levels and hospital admissions for peptic ulcer stratified by temperature (Table 3). Increases in the IQRs of all pollutants (PM$_{10}$, PM$_{2.5}$, SO$_2$, NO$_2$, CO, and O$_3$) were significantly associated with increased risk of hospital admissions on warm days (>23 °C). However, on cool days (<23 °C), only increases in PM$_{10}$, NO$_2$, and O$_3$ IQRs were found to be significantly associated with increased risk of hospital admission.

**Table 2.** Pearson correlation coefficients among air pollutants.

| Variable | PM$_{2.5}$ | PM$_{10}$ | NO$_2$ | CO | O$_3$ | SO$_2$ |
|----------|------------|-----------|--------|----|------|-------|
| PM$_{2.5}$ | 1.0        | 0.78      | 0.57   | 0.57| 0.28 | 0.64  |
| PM$_{10}$  | 1.0        | 1.00      | 0.33   | 0.36| 0.25 | 0.42  |
| NO$_2$     | 1.0        | 1.00      | 0.90   | −0.01| 0.56 | −0.01 |
| CO         | 1.0        | 1.00      | 0.90   | −0.01| 0.56 | −0.01 |
| O$_3$      | 1.0        | 1.00      | 1.16   | 1.16| 1.07 | 1.07  |
| SO$_2$     | 1.0        | 1.00      | 1.00   | 1.17| 1.07 | 1.07  |

**Table 3.** Association between exposure to air pollutant and hospital admissions for peptic ulcer in a single pollutant model for Taipei, Taiwan, 2009–2013.

| Pollutant | Whole period (n = 23,205) | >23 °C (1001 days) (n = 12,101) | <23 °C (825 days) (n = 11,104) |
|-----------|--------------------------|---------------------------------|-------------------------------|
| PM$_{10}$ | OR (95% CI) a,b           | OR (95% CI) a,b                 | OR (95% CI) a,b               |
| PM$_{2.5}$| 1.00 (0.98–1.02)          | 1.05 (1.01–1.08) c              | 1.04 (1.02–1.07) c            |
| SO$_2$    | 1.00 (0.98–1.02)          | 1.14 (1.09–1.18) c              | 1.10 (0.98–1.14) c            |
| NO$_2$    | 1.00 (0.98–1.02)          | 1.04 (1.00–1.08) c              | 0.92 (0.89–0.95)              |
| CO        | 1.00 (0.98–1.02)          | 1.16 (1.12–1.20) c              | 1.07 (1.04–1.11) c            |
| O$_3$     | 1.00 (0.98–1.02)          | 1.17 (1.12–1.21) c              | 1.02 (0.99–1.05)              |

a Interquartile range (IQR): PM$_{10}$ (26.07 µg/m$^3$), PM$_{2.5}$ (16.1 µg/m$^3$), SO$_2$ (1.71 ppb), NO$_2$ (8.18 ppb), CO (0.27 ppm), and O$_3$ (12.83 ppb); b control for temperature and humidity; c p < 0.05.
We used models including two pollutants to determine the independent effect of individual pollutants on hospitalizations for peptic ulcer (Table 4). This was performed to determine whether the effects of a particular pollutant would remain significant after each of the other pollutants was included in the model. On warm days, exposure to NO\(_2\) and O\(_3\) remained significantly associated with peptic ulcer admissions when combined with each of the other pollutants. PM\(_{10}\), NO\(_2\), and O\(_3\) remained significant in all two-pollutant models on cool days. The effect of SO\(_2\) remained significantly negative after incorporation of other pollutants into the model.

**Table 4.** Risk of peptic ulcer admissions per IQR increase in two-pollutant model by temperature \(^{a,b}\).

|                        | Control for PM\(_{10}\) | Control for PM\(_{2.5}\) | Control for SO\(_2\) | Control for NO\(_2\) | Control for CO | Control for O\(_3\) |
|------------------------|-------------------------|---------------------------|-----------------------|-----------------------|----------------|---------------------|
| **PM\(_{10}\)**        |                         |                           |                       |                       |                 |                     |
| >23 °C                 | –                       | –                         | 1.04 (1.00–1.08)       | 0.98 (0.94–1.02)      | 0.98 (0.94–1.02)| 1.02 (0.98–1.05)  |
| <23 °C                 | –                       | –                         | 1.09 (1.06–1.11)       | 1.03 (1.01–1.06)      | 1.05 (1.02–1.07)| 1.05 (1.03–1.07)  |
| **PM\(_{2.5}\)**       |                         |                           |                       |                       |                 |                     |
| >23 °C                 | –                       | –                         | 1.18 (1.12–1.24)       | 1.03 (0.97–1.09)      | 1.04 (0.99–1.10)| 1.10 (1.05–1.15)  |
| <23 °C                 | –                       | –                         | 1.13 (1.08–1.18)       | 0.96 (0.93–1.00)      | 0.99 (0.96–1.04)| 1.02 (0.99–1.06)  |
| **SO\(_2\)**          |                         |                           |                       |                       |                 |                     |
| >23 °C                 | 1.02 (0.98–1.06)        | 0.95 (0.91–0.99)          | –                     | 0.90 (0.86–0.95)      | 0.96 (0.92–0.99)| 1.01 (0.97–1.05)  |
| <23 °C                 | 0.87 (0.84–0.91)        | 0.84 (0.81–0.88)          | –                     | 0.81 (0.77–0.84)      | 0.84 (0.80–0.88)| 0.97 (0.93–1.00)  |
| **NO\(_2\)**          |                         |                           |                       |                       |                 |                     |
| >23 °C                 | 1.17 (1.12–1.22)        | 1.14 (1.07–1.19)          | 1.24 (1.18–1.30)       | –                     | 1.10 (1.02–1.18)| 1.14 (1.10–1.18)  |
| <23 °C                 | 1.06 (1.02–1.09)        | 1.09 (1.05–1.13)          | 1.22 (1.17–1.27)       | –                     | 1.28 (1.20–1.37)| 1.25 (1.20–1.30)  |
| **CO**                |                         |                           |                       |                       |                 |                     |
| >23 °C                 | 1.18 (1.13–1.23)        | 1.14 (1.09–1.20)          | 1.20 (1.14–1.25)       | 1.07 (0.99–1.16)      | –              | 1.16 (1.12–1.21)  |
| <23 °C                 | 0.99 (0.96–1.03)        | 1.02 (0.98–1.06)          | 1.14 (1.09–1.19)       | 0.82 (0.77–0.88)      | –              | 1.17 (1.12–1.21)  |
| **O\(_3\)**           |                         |                           |                       |                       |                 |                     |
| >23 °C                 | 1.10 (1.06–1.15)        | 1.06 (1.01–1.11)          | 1.11 (1.06–1.16)       | 1.07 (1.03–1.12)      | 1.10 (1.06–1.15)| –                   |
| <23 °C                 | 1.23 (1.18–1.29)        | 1.23 (1.18–1.29)          | 1.21 (1.15–1.26)       | 1.47 (1.39–1.55)      | 1.40 (1.32–1.48)| –                   |

\(^{a}\) IQR: PM\(_{10}\) (26.07 \(\mu g/m^3\)), PM\(_{2.5}\) (16.1 \(\mu g/m^3\)), SO\(_2\) (1.71 ppb), NO\(_2\) (8.18 ppb), CO (0.27 ppm), and O\(_3\) (12.83 ppb); \(^{b}\) Control for temperature and humidity.

### 4. Discussion

This study found significant positive relationships between ambient air NO\(_2\) and O\(_3\) levels and increased daily admissions for peptic ulcer on warm days and between PM\(_{10}\), NO\(_2\), and O\(_3\) and increased admissions on cool days. In this study, there was a significant negative effect of SO\(_2\) after the inclusion of other pollutants on cool days. We found no explanation for this in the literature, so it is possible that this inverse association was due to chance in the multiple significance testing process. It is also likely to be due to the collinearity between SO\(_2\) concentrations and other pollutant levels, which is a common problem in this type of study.

Investigations on the influence of air pollution on admissions for peptic ulcers are rare. Quan et al. \[20\] conducted a case-crossover study in Calgary and Edmonton, Canada. They reported that short-term increases in the level of ambient air pollutants did not increase the incidence of upper gastrointestinal bleeding secondary to peptic ulcer disease. This study only identified 2523 cases in adult residents. The small sample size might have resulted in low statistical power to detect an association. The study of Wong et al. \[21\], which followed 66,820 elderly persons in order to determine the effects of long-term exposure to PM\(_{2.5}\) on hospital admissions for peptic ulcer in Hong Kong, found the hazard ratio to be 1.18 (95% CI = 1.02–1.36%) for peptic ulcer hospitalization per each 10 \(\mu g/m^3\) increase in PM\(_{2.5}\). Tian et al \[22\] conducted a study of this association among the elderly population in Hong Kong. They reported a 7.6% (95% CI = 2.2–13.2%) increase in emergency admissions for peptic ulcer bleeding for each IQR increment (25.8 \(\mu g/m^3\)) in the 5-day moving average of NO\(_2\) concentration. Other pollutants (PM\(_{2.5}\), SO\(_2\), O\(_3\)) were not associated with peptic ulcer bleeding admissions in that...
study. In our current study of Taipei City, hospital admissions for peptic ulcer rose with increases in \( \text{NO}_2 \) and \( \text{O}_3 \) concentrations on warm days. We also found an association between \( \text{PM}_{10} \), \( \text{NO}_2 \), and \( \text{O}_3 \) and increases in daily peptic ulcer admissions on cool days. Bleeding is a common complication of peptic ulcer diseases. It is estimated that 40% of hospital admissions for peptic ulcer were due to bleeding [22]. In this regard, we postulate that short-term elevation in air pollution might trigger peptic ulcer bleeding events and increase the risk of hospital admissions for peptic ulcer, although the percentage of all hospitalizations for peptic ulcer that were due to bleeding is unknown in this study.

In this study, effects were observed on both warm and cool days, but they were larger on warm days (effect modification). The observed variation in effect estimates could be explained by variation in exposure patterns. People in Taipei are more likely to go outdoors and open the windows on warm days than on cool days (higher exposure); thus, monitored air pollutant levels may be closer to personal exposure on the warm days than on the cool days (better exposure assessment). This fact may reduce the air pollution effect on the cool days. On the other hand, differences in air pollution mixture between warm and cool days may also affect the effect estimates. Nevertheless, the degree to which variability and patterns of temperature may impact air pollution health effects needs to be further investigated.

Air pollutants are well known to directly affect the respiratory and circulation systems, but we can only hypothesize about possible pathophysiological processes underlying the association between short-term exposure and the development of and admissions for peptic ulcers. Studies have suggested that air pollutants can enter the aerodigestive tract via mucociliary clearance of PM from the lungs and can also enter that tract via ingestion of contaminated food and water [45–47]. Those studies also have suggested that the inhaled air pollutants may have direct toxic effects on epithelial cells after they have been transported to that location [45–47]. Still another explanation may be related to the fact that air pollutants increase intestinal permeability, alter the gut microbiota, and promote inflammation, thus contributing towards the development of peptic ulcer diseases in these ways [28–31]. Colonic microflora and elevated interleukin (IL)-8 and IL-17 levels in the small and large intestines have been found to be altered in mice guts exposed to PM [28]. In humans, exposure to \( \text{O}_3 \) can stimulate the production of tumor necrosis factor, IL-6, and IL-8 and induce systemic pro-inflammatory responses [48,49]. Such inflammatory responses, caused by disturbances in microbe composition in the gut and metabolic processing, have been found in mice fed PM [31]. Thus, exposure to air pollution, regardless of whether it is inhaled or ingested orally, may cause an imbalance in gut microbiota, thereby inducing an inflammatory response, which can lead to peptic ulcer. In addition, gaseous pollutants have been related to systemic inflammation, which may further induce or worsen adverse effects in the gastrointestinal tracts [50].

A case-crossover study design is a good approach to determining the health risk prompted by acute and intermittent exposures to possibly toxic substances. This design makes it possible to control for many confounders without the need for additional statistical modelling. All comparisons are within-individuals, so the design reduces confounding by time-invariant individual variables, including gender, age, underlying chronic diseases, and \( \text{Helicobacter pylori} \) infection. In addition, the time stratification approach we used in this study has been found to effectively control for seasonality, time trends, and chronic diseases and confounders that vary slowly over time [39,40].

We determined levels of exposure to air pollutants based on averaged data collected by fixed monitoring stations, which do not capture personal levels of exposure. The levels across various stations were averaged to access the population exposure levels to air pollution assuming that exposure was homogeneous throughout the area studied. This may raise some concerns over our estimates, especially since concentrations may differ between monitoring stations [51]. There may have been some errors in the measurement of exposure due to differences between population-averaged exposure and ambient air pollutant levels. This lack of individual measurements may have led to some misclassification and non-differential bias. However, this would produce an effect close to the null, resulting in an underestimation of the association [43,52].
There are some limitations with our study design. One limitation is that it is limited to one subtropical city. Another is that the peptic ulcer patients in this study include both the acute and the chronic types as well as both those with and without bleeding—patients presenting with different types of peptic ulcers and/or complications associated with peptic ulcers may have differences in symptom onset and/or delays of presentation to hospital [20,22]. This misclassification, however, was likely to be non-differential, which would tend to underestimate rather than overestimate the association. Another limitation is that our population belonged to the same race and ethnic background. Both of these limitations may reduce the possibility of generalizing our findings to other study areas. Another is that we did not take into account lifestyle behaviors, including how the use of air conditioning or time spent outdoors may affect personal exposures, which may vary widely with those of other study populations and locations. The design used in this study could have some problems with confounding by time-dependent covariates, such as the use of NSAIDs or aspirin. However, since the control days were the same day of the week in the same month, the differences in NSAIDs usage would likely not differ much over a short time period within the month studied [22]. Additionally, it is unlikely that day-to-day variation in NSAIDs usage would be correlated with the different levels of air pollutants.

5. Conclusions

In summary, in Taipei, Taiwan, short-term exposure to certain air pollutants (PM$_{10}$, NO$_2$, and O$_3$) is significantly associated with increased risk of hospital admissions due to peptic ulcers.

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Conflicts of Interest: The authors declare no conflict of interest.

References

1. Brunekreef, B.; Holgate, S.T. Air pollution and health. Lancet 2002, 360, 1233–1242. [CrossRef]
2. Adar, S.D.; Filigrana, P.A.; Clements, N.; Peel, J.L. Ambient coarse particulate matter and human health: A systematic review and meta-analysis. Curr. Environ. Health Rep. 2014, 8, 258–274. [CrossRef] [PubMed]
3. Atkinson, R.W.; Kang, S.; Anderson, H.R.; Mills, I.C.; Walton, H.A. Epidemiological time series studies of PM$_{2.5}$ and daily mortality and hospital admissions: A systematic review and meta-analysis. Thorax 2014, 69, 660–665. [CrossRef] [PubMed]
4. Shah, A.S.; Langrish, J.P.; Nair, H.; McAllister, D.A.; Hunter, A.L.; Donaldson, K.; Newby, D.E.; Mills, N.L. Global association of air pollution and heart failure: A systematic review and meta-analysis. Lancet 2013, 382, 1039–1048. [CrossRef] [PubMed]
5. Moore, E.; Chatzidiakou, L.; Kuku, M.O.; Jones, R.L.; Smeeth, L.; Beevers, S.; Kelley, F.J.; Barratt, B.; Quint, J.K. Global associations between air pollutants and chronic obstructive pulmonary disease hospitalizations: A systematic review. Ann. Am. Thorac. Soc. 2016, 13, 1814–1827. [CrossRef] [PubMed]
6. Brunekreef, B. Air pollution and life expectancy: Is there a relation? Occup. Environ. Health 1997, 54, 781–784. [CrossRef] [PubMed]
7. Pope, C.A.; Ezzati, M.; Dockery, D.W. Fine-particulate air pollution and life expectancy in the United States. N. Engl. J. Med. 2009, 360, 376–386. [CrossRef]
8. Ebenstein, A.; Fan, M.; Greenstone, M.; He, G.; Zhou, M. New evidence on the impact of sustained exposure to air pollution on life expectancy from China’s Huai river policy. Proc. Natl. Acad. Sci. USA 2017, 114, 10384–10389. [CrossRef]
9. Lipsett, M.; Hurley, S.; Ostro, B. Air pollution and emergency room visits for asthma in Santa Clara County, California. *Environ. Health Perspect.* 1997, 105, 216–222.

10. Hinwood, A.; De Klerk, N.; Rodriguez, C.; Jacoby, P.; Runnion, T.; Rye, P.; Landau, L.; Murray, F.; Feldwick, M.; Spickett, J. The relationship between changes in daily air pollution and hospitalizations in Perth, Australia 1992–1998: A case-crossover study. *Int. J. Environ. Health Res.* 2006, 16, 27–46.

11. Ballester, F.; Tenias, J.M.; Perez-Hoyos, S. Air pollution and emergency hospital admissions for cardiovascular diseases in Valencia, Spain. *J. Epidemiol. Commun. Health* 2001, 55, 57–65. [CrossRef] [PubMed]

12. Zmirou, D.; Barumandzadeh, T.; Balducci, F.; Ritter, P.; Laham, G.; Ghilardi, J.P. Short term effects of air pollution on mortality in the city of Lyon, France, 1985–1990. *J. Epidemiol. Commun. Health* 1996, 50 (Suppl. 1), S30–S35. [CrossRef] [PubMed]

13. Xu, C.; Kan, H.D.; Fan, Y.N.; Chen, R.J.; Liu, J.H.; Li, Y.F.; Zhang, Y.; Ji, A.L.; Cai, T.J. Acute effects of air pollution on enteritis admissions in Xi’an, China. *J. Toxicol. Environ. Health A* 2016, 79, 1183–1189. [CrossRef] [PubMed]

14. Kaplan, G.G.; Dixon, E.; Panaccione, R.; Fong, A.; Chen, L.; Szyszkwicz, M.; Wheeler, A.; MacLean, A.; Buie, W.D.; Leung, T.; et al. Effect of ambient air pollution on the incidence of appendicitis. *Can. Med. Assoc. J.* 2009, 181, 591–597. [CrossRef]

15. Kaplan, G.G.; Tanyingoh, D.; Dixon, E.; Johnson, M.; Wheeler, A.; Myers, R.P.; Bertazzon, S.; Saini, V.; Madsen, K.; Ghosh, S.; et al. Ambient ozone concentrations and the risk of perforated and nonperforated appendicitis: A multicity case-crossover study. *Environ. Health Perspect.* 2013, 131, 939–943. [CrossRef]

16. Chen, C.C.; Yang, C.Y. Effects of ambient air pollution exposure on frequency of hospital admissions for appendicitis in Taipei, Taiwan. *J. Toxicol. Environ. Health A* 2018, 81, 854–860. [CrossRef]

17. Kaplan, G.G.; Hubbard, J.; Sands, B.E.; Panaccione, R.; Ghosh, S.; Wheeler, A.J.; Villeneuve, P.J. The inflammatory bowel diseases and air pollution: An association. *Am. J. Gastroenterol.* 2010, 105, 2412–2419. [CrossRef]

18. Ananthakrishnan, A.N.; McGinley, E.L.; Binion, D.G.; Saean, K. Ambient air pollution correlates with hospitalizations for inflammatory bowel diseases: An ecologic analysis. *Inflamm. Bowel Dis.* 2011, 17, 1138–1145. [CrossRef]

19. Kaplan, G.G.; Szyszkwicz, M.; Fichna, J.; Rowe, B.H.; Porada, E.; Vincent, R.; Madsen, K.; Ghosh, S.; Storr, M. Non-specific abdominal pain and air pollution: A novel association. *PLoS ONE* 2012, 7, e47669. [CrossRef]

20. Quan, S.; Yang, H.; Tanyingoh, D.; Villeneuve, P.J.; Stieb, D.M.; Johnson, M.; Hilsden, R.; Madsen, K.; van Zanten, S.V.; Novak, K.; et al. Upper gastrointestinal bleeding due to peptic ulcer disease is not associated with air pollution: A case-crossover study. *BMJ Gastroenterol.* 2015, 15, 131. [CrossRef]

21. Wong, C.M.; Tsang, H.; Lai, H.K.; Thach, T.Q.; Thomas, G.N.; Chan, K.P.; Lee, S.Y.; Ayres, J.G.; Lam, T.H.; Leung, W.K. STROBE-Long Term exposure to ambient fine particulate air pollution and hospitalization due to peptic ulcers. *Medicine* 2016, 95, e3543. [CrossRef]

22. Tian, L.; Qiu, H.; Sun, S.; Tsang, H.; Chan, K.P.; Leung, W.K. Association between emergency admissions for peptic ulcer bleeding and air pollution: A case-crossover analysis in Hong Kong’s elderly population. *Lancet Planet Health* 2017, 1, e74–e81. [CrossRef]

23. Laine, L. Upper gastrointestinal bleeding due to a peptic ulcer. *N. Engl. J. Med.* 2016, 374, 2367–2376. [CrossRef]

24. Sostres, C.; Carrera-Lasfuentes, P.; Benito, R.; Roncales, P.; Arruebo, M.; Arroyo, M.T.; Bujanda, L.; Garcia-Rodriguez, L.A.; Lanas, A. Peptic ulcer bleeding risk. The role of Helicobacter pylori infection in NSAID/low-dose aspirin users. *Am. J. Gastroenterol.* 2015, 110, 684–689. [CrossRef]

25. Yoon, H.; Kim, S.G.; Jung, H.C.; Song, I.S. High recurrence rate of idiopathic peptic ulcers in long-term follow-up. *Gut Liver* 2013, 7, 175–181. [CrossRef]

26. Lijima, K.; Kanno, T.; Koike, T.; Shimosogawa, T. Helicobacter pylori-negative, non-steroidal anti-inflammatory drug: Negative idiopathic ulcers in Asia. *World J. Gastroenterol.* 2014, 20, 706–713.

27. Nomura, T.; Ohkusa, T.; Araki, A.; Chuganji, Y.; Momoi, M.; Takashimizu, I.; Watanabe, M. Influence of climatic factors in the incidence of upper gastrointestinal bleeding. *J. Gastroenterol. Hepatol.* 2001, 16, 619–623. [CrossRef]

28. Kish, L.; Hotte, N.; Kaplan, G.G.; Vincent, R.; Tso, R.; Ganzle, M.; Rioux, K.P.; Thiesen, A.; Barkema, H.W.; Wine, E.; et al. Environmental particulate matter induces murine intestinal inflammatory response and alters the gut microbiome. *PLoS ONE* 2013, 8, e62220. [CrossRef]
29. Mutlu, E.A.; Engen, P.A.; Soberanes, S.; Urich, D.; Forsyth, C.B.; Nigdelioglu, R.; Chiarella, S.E.; Radigan, K.A.; Gonzalez, A.; Jakate, S.; et al. Particulate matter air pollution causes oxidant-mediated increase in gut permeability in mice. Part. Fibre Toxicol. 2011, 8, 19. [CrossRef]

30. Salim, S.Y.; Jovel, J.; Wine, E.; Kaplan, G.G.; Vincent, R.; Thiesen, A.; Barkema, H.W.; Madsen, K.L. Exposure to ingested airborne particulate matter increases mucosal exposure to bacteria and induces early onset of inflammation in neonatal IL-10-deficient mice. Inflamm. Bowel Dis. 2014, 20, 1129–1138. [CrossRef]

31. Salim, S.Y.; Kaplan, G.G.; Madsen, K.L. Air pollution effects on the gut microbiota: A link between exposure and inflammatory disease. Gut Microbes 2014, 5, 215–219. [CrossRef]

32. Chang, C.C.; Tsai, S.S.; Ho, S.C.; Yang, C.Y. Air pollution and hospital admissions for cardiovascular disease in Taipei, Taiwan. Environ. Res. 2005, 98, 114–119. [CrossRef] [PubMed]

33. Kuo, H.W.; Tsai, S.S.; Tiao, M.M.; Yang, C.Y. Epidemiologic features of chronic kidney disease in Taiwan. Am. J. Kidney Dis. 2007, 49, 46–55. [CrossRef]

34. Tsai, S.Y.; Jovel, J.; Wine, E.; Kaplan, G.G.; Vincent, R.; Thiesen, A.; Barkema, H.W.; Madsen, K.L. Exposure to ingested airborne particulate matter increases mucosal exposure to bacteria and induces early onset of inflammation in neonatal IL-10-deficient mice. Inflamm. Bowel Dis. 2014, 20, 1129–1138. [CrossRef]

35. Maclure, M. The case-crossover design: A method for studying transient effects on the risk of acute events. Am. J. Epidemiol. 1991, 133, 144–153. [CrossRef]

36. Marshall, R.J.; Jackson, R.T. Analysis of case-crossover designs. Stat. Med. 1993, 12, 2333–2341. [CrossRef]

37. Mittleman, M.A. Optimal referent selection strategies in case-crossover studies: A settled issue. Epidemiology 1999, 10, 567–575. [CrossRef]

38. Levy, D.; Lumley, T.; Kaufman, J.; Checkoway, H. Referent selection in case-crossover analyses of acute health effects of air pollution. Epidemiology 2001, 12, 186–192. [CrossRef] [PubMed]

39. Janes, H.; Sheppard, L.; Lumley, T. Case-crossover analyses of air pollution exposure data: Referent selection strategies and their implications for bias. Epidemiology 2005, 16, 717–726. [CrossRef]

40. Mittleman, M.A. Optimal referent selection strategies in case-crossover studies: A settled issue. Epidemiology 2005, 16, 15–16. [CrossRef]

41. Lu, Y.; Zeger, S.L. On the equivalence of case-crossover and time series methods in environmental epidemiology. Biostatistics 2007, 8, 337–344. [CrossRef]

42. Neas, L.M.; Schwartz, J.; Dockery, D. A case-crossover analysis of air pollution and mortality in Philadelphia. Environ. Health Perspect. 1999, 107, 629–631. [CrossRef] [PubMed]

43. Katsouyanni, K.; Touloumi, G.; Spix, C.; Schwartz, J.; Balducci, F.; Medina, S.; Rossi, G.; Wojtyniak, B.; Sunyer, J.; Bacharova, L.; et al. Short term effects of ambient sulphur dioxide and particulate matter on mortality in 12 European cities: Results from time series data from the APHEA project. Br. Med. J. 1997, 314, 1658–1663. [CrossRef] [PubMed]

44. Dominici, F.; Peng, R.D.; Bell, M.L.; Pham, L.; McDermott, A.; Zeger, S.L.; Samet, J.M. Fine particulate air pollution and hospital admission for cardiovascular and respiratory diseases. J. Am. Med. Assoc. 2006, 295, 1127–1134. [CrossRef]

45. Beamish, L.A.; Osornio-Vargas, A.R.; Wine, E. Air pollution: An environmental factor contributing to intestinal disease. J. Crohn’s Colit. 2011, 5, 279–286. [CrossRef]

46. Moller, W.; Haussinger, K.; Winkler-Heil, R.; Stahlhofen, W.; Meyer, T.; Heyder, J. Mucociliary and long-term particle clearance in the airways of healthy nonsmoker subjects. J. Appl. Physiol. 2004, 97, 2200–2206. [CrossRef]

47. Kreyling, W.G.; Blanchard, J.D.; Godleski, J.J.; Haeussermann, S.; Heyder, J.; Hutzler, P.; Schulz, H.; Sweeney, T.D.; Takenaka, S.; Siesenis, A. Anatomic localization of 24- and 96-h particle retention in canine airways. J. Appl. Physiol. 1999, 87, 269–284. [CrossRef]

48. Bossen, J.; Pourazar, J.; Forsberg, B.; Adelroth, E.; Sandstrom, T.; Blomberg, A. Ozone enhances the airway inflammation initiated by diesel exhaust. Respir. Med. 2007, 101, 1140–1146. [CrossRef] [PubMed]

49. Thompson, A.M.; Zanobetti, A.; Silverman, F.; Schwartz, J.; Coull, B.; Uresh, B.; Speck, M.; Brook, J.R.; Manno, M.; Gold, D.R. Baseline repeated measures from controlled human exposure studies: Associations between ambient air pollution exposure and the systemic inflammatory biomarkers IL-6 and fibrinogen. Environ. Health Perspect. 2010, 118, 120–124. [CrossRef]
50. Calderon-Garciduenas, L.; Franco-Lira, M.; Torres-Jardon, R.; Henriquez-Roldan, C.; Barragan-Mejia, G.; Valencia-Salazar, G.; Gonzalez-Maciel, A.; Reynoso-Robles, R.; Villarreal-Calderon, R. Pediatric respiratory and systemic effects of chronic air pollution exposure: Nose, lung, heart, and brain pathology. *Toxicol. Pathol.* 2007, 35, 154–162. [CrossRef]

51. Samet, J.M.; Zeger, S.L.; Dominici, F.; Curriero, F.; Coursac, I.; Dockery, D.W.; Schwartz, J.; Zanobetti, A. The National Morbidity, Mortality, and air pollution study. Part II. Morbidity and Mortality from air pollution in the United States. *Res. Rep. Health Eff. Inst.* 2000, 94, 5–79.

52. Zeger, S.L.; Thomas, D.; Dominici, F.; Samet, J.M.; Schwartz, J.; Dockery, D.; Cohen, A. Exposure measurement error in time-series studies of air pollution: Concepts and consequences. *Environ. Health Perspect.* 2000, 108, 419–426. [CrossRef]