Research Article

Assessment of Impacts Produced by Anthropogenic Sources in a Little City near an Important Industrial Area (Modugno, Southern Italy)

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An annual monitoring campaign of VOCs, consisting of twelve sampling periods, was carried out from June 2008 to June 2009 in Modugno, a city located in the Apulia region (Southern Italy), in order to assess the urban air quality, identify the main emission sources, and quantify the cancer and non-cancer risk attributable to inhalation exposures. Monitoring, carried out by using the Radiello diffusive samplers, was conducted in eleven sampling sites throughout the city taking into account the traffic density and the architecture of the city. From the study of the data, it was found that, among all considered VOCs, benzene, toluene, ethylbenzene, and xylenes (BTEX) are the pollutants at higher concentration. The analysis of VOC concentrations, the study of the topography of the city, and the use of different diagnostic ratios between the BTEX species showed that the vehicular traffic emissions were the predominant source of VOCs in the urban area of Modugno. Despite that the annual concentration of benzene is lower than the regulatory limit, the estimation of cancer risk showed that the global lifetime cancer risk attributed to the investigated VOC exposure was not negligible and therefore should be taken into account in future regulatory approaches.

1. Introduction

Because of their considerable impact on the environment and human health, volatile organic compounds (VOCs) are considered essential parameters for assessing the air quality in indoor and outdoor environments. A large number of VOCs are usually emitted into the atmosphere of industrialized countries as discussed in different studies [1–4]. VOCs play a significant role in the formation of oxidants like ozone and peroxyacetyl nitrate (PAN) in the troposphere and secondary aerosols [5]. Although motor vehicles are the major source of VOCs in urban areas, these pollutants are also emitted from chemical plants, petroleum refineries, dry cleaning establishments, filling stations, painting operations, and even many household products [6]. Benzene, toluene, ethylbenzene, and xylenes (BTEX) are generally compounds associated with traffic emissions; toluene is also released with the use of solvents (painting, printing, dry cleaning, etc.) as discussed in many works [7–9]. On a global scale, the biogenic VOC emissions, mainly isoprene, α-pinene and limonene, dominate over the anthropogenic sources [10, 11]. In and around urban areas, anthropogenic emissions of VOCs are usually more significant [12–15]. The chemical diversity of the VOC group is reflected in the diversity of the health effects that individual VOCs can cause, ranging from relatively inert VOCs with no known health effects to the highly toxic effects of reactive VOCs [4, 16, 17]. In general, chronic health effects of VOCs can be classified as either non-carcinogenic or carcinogenic. The main noncarcinogenic chronic effects are irritative, sensory effects, damage to the liver, kidneys, and central nervous system, asthma, and other respiratory effects. The main carcinogenic effects are lung, blood (leukaemia and non-Hodgkin's lymphoma), liver, kidney, and biliary tract cancer. Most researches have focused on the urban levels
of VOCs, especially aromatic and chlorinated organic compounds, due to the known and suspected carcinogenic nature of these species [4, 16, 17]. In fact, several VOCs found in urban air are accepted carcinogens (1,3-butadiene, benzene, formaldehyde, and acetaldehyde). The Organisation for Economic Co-operation and Development (OECD) classified these compounds according to their cancer risk. Benzene and 1,3-butadiene account for 68% of the cancer risk from all vehicle related pollutants, whereas in comparison, particulate matter accounts for only 28% [18]. Benzene has been shown to cause cancer in both animals and humans; therefore, it is currently classified by the Environmental Protection Agency (EPA), the American Conference of Governmental Industrial Hygienist (ACGIH), and the International Agency for Research on Cancer (IARC) as a human carcinogen [19–21]. In Europe countries, only the ambient concentration of benzene, limited to an annual average of 5 \( \mu \text{g/m}^3 \), is regulated [22]. Other alkylbenzenes cannot be classified as a carcinogen compounds, since there is inadequate evidence for them carcinogenicity in humans even if many studies have shown the several effects linked to the exposure to these pollutants [23–27].

In the present study, an annual monitoring campaign of VOCs was performed in eleven sampling sites of Modugno, a town located in Apulia region (southern Italy), in order to assess the urban air quality, identify the main emission sources, and quantify the cancer and no-cancer risk attributable to inhalation exposures. Moreover, a comparison of the monitored VOC concentrations with the levels detected in other Apulia cities and other urban areas was conducted.

2. Materials and Methods

2.1. Sampling Sites. An annual monitoring campaign of VOCs was carried out from June 2008 to June 2009 in Modugno, a town located in Apulia region (Southern Italy). The samples were collected in eleven sampling sites throughout the city for a week at month (Figure 1) to make a detailed characterization of the composition of Modugno urban air.

Since the 1970s, this town was mainly dedicated to agriculture but after the construction of the industrial zone, it has become one of the most important industrial sites of the region. The industrial area of Modugno is characterized by the presence of national and international enterprises dealing with mechanical, transport, logistic, chemical, and pharmaceutical activities; close to the city, there is also a natural gas energy plant production. The selection of the sites was planned taking into account the traffic density and the architecture of the city: the sites were positioned in order to be as representative as possible of the mean concentration levels of pollutants for each area. In particular, the sampling sites were divided among hot-spot points (4 locations mainly associated with main roads where high traffic density was identified) and 7 points of background indicative of the levels of exposure of not near to an intense circulation of traffic.

2.2. Sampling Method. VOCs were sampled with Radiello diffusive samplers (Fondazione Salvatore Maugeri, Padova, Italy) suitable for thermal desorption. The sampling system is made up of a cylindrical adsorbing cartridge housed coaxially inside a cylindrical diffusive body of polycarbonate and microporous polyethylene. The cartridges are composed of a cylindrical stainless steel net (100 mesh) with the external diameter of 4.8 mm, containing 350 mg of 35–50 mesh of Carbograph 4. Before the sampling, the cartridges were conditioned and analysed to verify the blank levels [28]. Each sampler was exposed and after sampling it was sealed in a sealed glass tube and brought to the laboratory for the analysis. Contemporarily the main meteorological parameters (ambient temperature, wind, atmospheric pressure, and natural radioactivity) were monitored.

2.3. Analytical Method. The analyses were carried out by using a thermal desorber (Markes International Ltd., Unity) equipped with an autosampler [Markes mod. ULTRA TD]
Table 1: Annual mean concentrations of the investigated pollutants in each sampling sites expressed as μg/m³.

| Step                  | Parameter                        | Value          |
|-----------------------|----------------------------------|----------------|
| Cartridge conditioning| Conditioning time                | 15 min         |
|                       | Conditioning temperature          | 310°C          |
|                       | Gas flow                         | 1 mL/min       |
|                       | Split flow                        | 50 mL/min      |
| Adsorbent tube desorption| Purge time                      | 1 min—trap in line |
|                       | Desorption time                   | 10 min         |
|                       | Desorption temperature            | 300°C          |
|                       | Temperature of cold trap          | −10°C          |
|                       | Desorption flow                   | 20 mL/min      |
| Focusing trap desorption| Desorption time                  | 3 min          |
|                       | Temperature of cold trap desorption| 300°C        |
|                       | Split flow                        | 44 mL/min      |
|                       | Temperature transfer line         | 150°C          |
| GC analysis           | Gas carrier                       | He             |
|                       | Gas flow                          | 1.7 mL/min     |
|                       | Analytical column                 | Polyethylene glycol 30 m × 0.25 mm ID, 0.25μm |
|                       | SUPELCOWAX (Supelco)              | Stationary phase thickness  |
|                       | Oven temperature                  | 80°C per 1 min, 20°C/min fino a 270°C, 270°C per 3 min. |

provided with 100 positions and coupled with a gas chromatograph (Agilent GC-6890 PLUS) and a mass selective detector (Agilent MS-5973 N). The thermal desorber provides a two-stage mechanism: in the former the analytes are desorbed from the sample tube and refocused into a cold trap; in the latter they are desorbed from the trap and carried into the GC column [28]. The parameters of thermal desorption and GC-MS analysis are listed in Table 1. The standard solutions were prepared by successive dilution in methanol of a VOC standard mixture at 2000 μg/mL (Ultra scientific Cas-5997). To quantify the samples, the calibration curves were prepared injecting 1 μL of the standard solutions into a tube; the spiked adsorbent tubes were then thermally desorbed in the same conditions of time, gas flow, and split ratio of the samples. The sampling rates, Q values supplied by manufacturer, were useful to calculate the real concentration of compound in the atmosphere (C) by GC quantification of analytes mass, m. Q is function of the diffusive coefficient D, which is a thermodynamic property of each chemical substance. The sampling rate has the dimensions of a gaseous flow: if m is expressed in μg, the sampling period in minutes and C in μg·L⁻¹, Q is expressed in L·min⁻¹.

3. Results and Discussions

Annual mean concentrations (μg/m³) of each monitored compound were calculated in all investigated sites and are listed in Table 2. The sum of the detected compounds (VOC-Sum) and sum of benzene, toluene ethylbenzene, and xylenes (BTEX) in each sampling site are also reported. From Table 1, it can be observed that, among all considered VOCs, BTEX are the pollutants at higher concentration in all sites. Mean BTEX/VOC-Sum ratio percentage over the city was equal to 63 and the highest ratios were obtained for sites 1, 5, and 8.

This finding is also evident in Figure 2 where the trend of VOC-Sum, differentiated among BTEX and other compounds, is shown. Nevertheless it was found that the annual benzene concentrations in urban area of Modugno was lower than the limit value for 2009 (6 μg/m³) [22]. The study of the topography of the area revealed that site 5 is located in a narrow road with heavy traffic and low speed of travel due to the presence of a level crossing. Moreover, the site is impacted by emissions produced by the main artery of the city characterized by heavy traffic. Site 8 represents a critical area of the city center as it is located in a very busy road with many intersections and traffic lights that slow down the speed of travel. Site 1 was near one of the main access roads to the city with high traffic density. The highest concentrations of VOCs were monitored in August, October, January, and June (see Figure 3).

Furthermore, high concentrations of decane (1.5–8.0 μg/m³) were revealed in most sites in February. Sites 3 and 4, located in a peripheral area of the city and near the industrial area, were characterized by the lowest levels of VOCs and BTEX. It is known that toluene and benzene concentrations are reduced in atmosphere through their reaction with OH radicals with the rate of toluene approximately 5 times larger than that of benzene. Therefore, ambient T/B ratios that are significantly lower than vehicular emission ratios are expected to have travelled and degraded,
Figure 3: Trend of VOC Sum, differentiated among aromatics aliphatics and other compounds, in the most polluted sites.

Figure 4: BTEX concentrations during the warm season (April–September) and cold season (October to March).

Figure 5: Correlation analysis among BTEX concentrations in all investigated sites.

whereas higher T/B ratios may reflect relatively fresh vehicular emission sources. In the present study, annual values of T/B ratio in Modugno (2.1) were nearly similar to those found in urban areas of many Apulian cities such as in Bari (2.0) and Canosa di Puglia (3.4) (see Table 3) [29, 30].

Similar ratios were observed by other researches in urban sites: in Rome (2.8), Izmir (2.0), Santiago (2.01), Paris (2.9–3.4), and Copenhagen (2.2) [31–35]. The correlation analysis among annual VOC concentrations in all sites was also performed (Table 4). Significant positive correlation coefficients were found among BTEX concentrations (≥0.99) and among aromatic hydrocarbons with the exception of styrene. This finding is indicative that local traffic emissions were large contributors to these compounds.

Several studies identified a group of ratios used as an indicator of the photochemical age of the air mass. In addition
Table 2: Annual mean concentrations of the investigated pollutants in each sampling site expressed as $\mu g/m^3$.

| Pollutant                  | sitio 1 $\mu g/m^3$ | sitio 2 $\mu g/m^3$ | sitio 3 $\mu g/m^3$ | sitio 4 $\mu g/m^3$ | sitio 5 $\mu g/m^3$ | sitio 6 $\mu g/m^3$ | sitio 7 $\mu g/m^3$ | sitio 8 $\mu g/m^3$ | sitio 9 $\mu g/m^3$ | sitio 10 $\mu g/m^3$ | sitio 11 $\mu g/m^3$ | SD |
|----------------------------|---------------------|---------------------|---------------------|---------------------|---------------------|---------------------|---------------------|---------------------|---------------------|---------------------|---------------------|-----|
| Benzene                   | 1.83                | 1.00                | 0.48                | 0.88                | 0.42                | 0.89                | 0.48                | 2.57                | 1.00                | 0.55                | 0.89                | 0.43 |
| Heptane                   | 1.03                | 0.40                | 0.61                | 0.32                | 0.49                | 0.22                | 0.90                | 1.51                | 0.53                | 0.69                | 1.01                | 0.27 |
| Toluene                   | 4.22                | 1.01                | 1.79                | 0.72                | 1.34                | 0.44                | 1.63                | 0.55                | 6.27                | 1.59                | 2.50                | 0.75 |
| Tetrachloroethylene       | 0.38                | 0.10                | 0.30                | 0.08                | 0.27                | 0.07                | 0.29                | 0.10                | 0.35                | 0.11                | 0.29                | 0.09 |
| Butyl acetate             | 0.33                | 0.18                | 0.25                | 0.11                | 0.27                | 0.14                | 0.28                | 0.15                | 0.34                | 0.13                | 0.20                | 0.09 |
| Ethylbenzene              | 0.95                | 0.26                | 0.42                | 0.18                | 0.33                | 0.12                | 0.38                | 0.15                | 1.41                | 0.36                | 0.57                | 0.14 |
| m-, p-Xylenes             | 3.00                | 0.71                | 1.14                | 0.53                | 0.79                | 0.26                | 0.94                | 0.31                | 4.70                | 1.70                | 1.71                | 0.46 |
| o-Xylene                  | 1.18                | 0.30                | 0.43                | 0.18                | 0.30                | 0.10                | 0.34                | 0.10                | 1.87                | 0.46                | 0.68                | 0.21 |
| Styrene                   | 0.20                | 0.07                | 0.39                | 0.12                | 0.16                | 0.11                | 0.31                | 0.31                | 0.23                | 0.04                | 0.17                | 0.05 |
| a-Pinene                  | 0.69                | 1.35                | 0.33                | 0.13                | 0.42                | 0.39                | 0.55                | 0.46                | 0.43                | 0.56                | 0.36                | 0.23 |
| 1,3,5-Trimethylbenzene     | 0.28                | 0.12                | 0.34                | 0.07                | 0.10                | 0.05                | 0.10                | 0.05                | 0.54                | 0.26                | 0.21                | 0.15 |
| 2-Ethyl-1-hexanol          | 0.11                | 0.19                | 0.18                | 0.22                | 0.26                | 0.36                | 0.19                | 0.24                | 0.37                | 0.65                | 0.18                | 0.32 |
| Decane                    | 0.34                | 0.45                | 1.31                | 2.42                | 1.00                | 1.95                | 0.24                | 0.13                | 0.94                | 2.12                | 0.31                | 0.25 |
| 1,4-Dichlorobenzene       | 0.07                | 0.03                | 0.06                | 0.04                | 0.06                | 0.03                | 0.06                | 0.03                | 0.07                | 0.03                | 0.06                | 0.03 |
| 1,2,4-Trimethylbenzene     | 0.47                | 0.44                | 0.21                | 0.20                | 0.15                | 0.11                | 0.16                | 0.13                | 1.08                | 0.74                | 0.33                | 0.20 |
| Limonene                  | 0.19                | 0.27                | 0.19                | 0.13                | 0.14                | 0.12                | 0.23                | 0.18                | 0.24                | 0.33                | 0.17                | 0.19 |
| 1,2,3-Trimethylbenzene     | 0.32                | 0.36                | 0.16                | 0.17                | 0.15                | 0.18                | 0.12                | 0.12                | 0.51                | 0.40                | 0.27                | 0.32 |
| VOCSum                    | 15.59               | 6.96                | 8.70                | 6.10                | 7.09                | 5.07                | 7.63                | 4.99                | 23.45               | 10.56               | 10.25               | 5.30 |
| BTEX                      | 11.18               | 2.99                | 4.78                | 2.09                | 3.63                | 1.34                | 4.18                | 1.59                | 16.81               | 4.50                | 6.67                | 2.11 |
| Percentage (BTEX/VOCSum)  | 72                  | 55                  | 51                  | 55                  | 72                  | 65                  | 56                  | 71                  | 69                  | 66                  | 66                  |     |
of xylenes were found in sites 3 and 4 (X/E attributable to traffic emissions (3:4:1:4) while low ratios by one of the emission sources [39]. In most sites, benzene sites across a city to identify areas influenced more strongly (T/B), 2.988 ((m + p)/E), 1.45 ((m + p)/B), and 0.56 (O/B). confirming the presence of local emission sources: 2.13 study were in agreement with those found by Khoder [38] and 0.85 (O/B) [38]. The average ratios found in the present as follows: 2.45 (T/B), 3.28 ((m + p)/E), 1.61 ((m + p)/B), and 0.85 (O/B) [38]. The average ratios found in the present study were in agreement with those found by Khoder [38] confirming the presence of local emission sources: 2.13 (T/B), 2.988 ((m + p)/E), 1.45 ((m + p)/B), and 0.56 (O/B). Furthermore, B : T : E : X ratios can be useful for comparing sites across a city to identify areas influenced more strongly by one of the emission sources [39]. In most sites, benzene was in lower concentration ratios than would normally be attributable to traffic emissions (3 : 4 : 1 : 4) while lower ratios of xylenes were found in sites 3 and 4 (X/E = 23) located near the industrial area. Analyzing BTEX concentrations during the warm season (April–September) and cold season (October to March), as shown in Figure 4, we can see that the sites are characterized by concentrations approximately constant during the year. BTEX concentrations did not show a seasonal trend. This finding can be explained considering that the high emissions of vehicular traffic, usually close to monitoring sites, contributed predominantly to the levels of VOCs monitored.

The correlation analysis among BTEX concentrations in all sites confirmed this hypothesis (Figure 5). A significant positive correlation was found among investigated sites with the exception of sites 6 and 7 characterized by lower correlation coefficients.

This study focused also on characterizing the risk of exposure to VOCs by means of inhalation in people living in the city. The estimates for the non-cancer risk assessment were based on the comparison of the ambient concentrations with their respective chronic noncancer inhalation level at which no adverse effects are expected for single VOCs. These levels are expressed as Reference Concentrations (RFCs)—used by the USEPA Integrated Risk Information System (IRIS); Minimum Risk Levels (MRLs)—used by the Agency for Toxic Substances and Disease Registry (ATSDR); or Reference Exposure Levels (RELs) used by the California Office of Environmental Health Hazard Assessment (OEHHA) [40–42]. Therefore, the hazard ratio (HR) of each compound was calculated by dividing its concentration (expressed in μg m⁻³) by its corresponding Reference Concentration (RFC, also in μg m⁻³) [24]. The calculated average HRs were less than 1 for all the studied VOCs, which means that their concentrations were commonly below the level of concern (see Table 5).

Cancer risk assessment was calculated for the carcinogenic VOCs detected in the samples, whose unit risk (UR) values were established by an official agency. These URs were extracted from databases provided by WHO, IRIS, and OEHHA giving priority to the WHO and IRIS URs, in that order (see Table 4) [23, 40, 42]. The lifetime cancer risk (LCR) attributable to inhalation exposures was calculated by multiplying each UR estimated value by the average concentrations (in μg m⁻³) of each compound. The UR of each compound corresponds to excess lifetime cancer risk calculated as a result of the continuous exposure to an agent at a concentration of 1 μg m⁻³ over a lifetime of 70 years [24, 40]. In this study, the individual cancer risk of each VOC was assumed to be additive, and therefore, the global lifetime cancer risk was considered as the sum of the LCRs of the individual compounds. In accordance with previous studies, compounds with an attributable cancer risk of over 10⁻⁴ can be considered as a “definite risk,” between 10⁻⁵ and 10⁻⁴ as a “probable risk” and between 10⁻⁵ and 10⁻⁶ as a “possible risk” [24]. Among the monitored compounds, benzene, toluene, tetrachloroethylene, and 1,4-dichlorobenzene are the compounds having the UR estimated value. For each compound, the possible cancer risk for humans was determined (mean LCRs ranged from 7.21 × 10⁻⁷ to 8.29 × 10⁻⁶). Moreover, the global lifetime cancer risk was calculated in each site summing the LCRs of the individual compounds (Table 6). Similar global lifetime cancer risks were found in the investigated sites: in each one, the estimated risk was higher than the USEPA guideline (10⁻⁶). This finding showed that the average cancer risk attributed to these VOC exposure was not negligible and therefore should be taken into account to preserve the health of citizens.

### 4. Conclusion

In this study, an annual monitoring campaign of VOCs was performed in eleven sampling sites of Modugno, a city located in Apulia region (southern Italy), in order to assess the urban air quality, identify the main emission sources, and quantify the cancer and no-cancer risk attributable to inhalation exposures. The analysis of VOC concentrations, the study of the topography of the city, and the use of different diagnostic ratios between the BTEX species allowed

### Table 3: Diagnostic ratios between the BTEX compounds.

| Sites   | T/B   | (m + p)/B | O/B   | (m + p)/E | B : T : E : X |
|---------|-------|-----------|-------|-----------|---------------|
| Site 1  | 2.30  | 1.64      | 0.64  | 3.16      | 2 : 4 : 1 : 4  |
| Site 2  | 1.79  | 1.14      | 0.43  | 2.69      | 2 : 4 : 1 : 4  |
| Site 3  | 1.52  | 0.89      | 0.34  | 2.40      | 2 : 4 : 1 : 3  |
| Site 4  | 1.82  | 1.05      | 0.38  | 2.48      | 2 : 4 : 1 : 3  |
| Site 5  | 2.44  | 1.83      | 0.73  | 3.32      | 2 : 4 : 1 : 4  |
| Site 6  | 2.06  | 1.41      | 0.56  | 3.01      | 2 : 4 : 1 : 4  |
| Site 7  | 2.12  | 1.54      | 0.58  | 3.16      | 2 : 4 : 1 : 4  |
| Site 8  | 2.49  | 1.93      | 0.76  | 3.32      | 2 : 4 : 1 : 4  |
| Site 9  | 2.48  | 1.61      | 0.61  | 3.03      | 2 : 4 : 1 : 4  |
| Site 10 | 2.01  | 1.46      | 0.57  | 3.11      | 2 : 4 : 1 : 4  |
| Site 11 | 2.38  | 1.45      | 0.56  | 3.09      | 2 : 4 : 1 : 4  |
| Mean    | 2.13  | 1.45      | 0.56  | 2.98      |               |
| RSD%    | 15    | 22        | 24    | 11        |               |
| Median  | 2.12  | 1.46      | 0.57  | 3.09      |               |
| Min     | 1.52  | 0.89      | 0.34  | 2.40      |               |
| Max     | 2.49  | 1.93      | 0.76  | 3.32      |               |
Table 4: Correlation coefficients among VOC concentrations in all investigated sites: coefficients with values greater than 0.5 are in bold.

|        | Benzene | Heptane | Toluene | Tetrachloroethylene | Butyl acetate | Ethylbenzene | m-, p-Xylenes | Styrene | a-Pinene | 1,3,5-Trimethilbenzene | 2-Ethyl-1-hexanol | Decane | 1,4-Dichlorobenzene | 1,2,4-Trimethylbenzene | Limonene | 1,2,3-Trimethylbenzene |
|--------|---------|---------|---------|---------------------|---------------|--------------|---------------|---------|----------|------------------------|-----------------|--------|----------------------|----------------------|----------|----------------------|
| Benzene| 1.00    | 0.91    | 0.99    | 0.51                | 0.62          | 0.99         | 0.99          | -0.04   | 0.01     | 0.95                   | 0.23            | 0.03   | 0.88                  | 0.95                  | 0.49     | 0.93                  |
| Heptane| 1.00    | 0.92    | 0.49    | 0.13                | 0.57          | 0.58         | -0.07         | -0.39   | 0.96     | -0.33                  | -0.27           | 0.67   | 0.58                  | 0.06                  | 0.71     | 0.95                  |
| Toluene| 1.00    | 0.60    | 0.91    | 0.12                | 0.00          | -0.04        | 0.96          | 0.18    | -0.05    | 0.87                   | 0.95            | 0.49   | 0.51                  | 0.49                  | 0.40     | 0.94                  |
| Tetrachloroethylene| 1.00 | 0.57| 0.58 | -0.07 | 0.96 | 0.02 | 0.07 | 0.07 | 0.11 | 0.04 | 0.85 | 1.00 | 0.05 | 1.00 | 0.05 | 1.00 | 0.05 |
| Butyl acetate| 1.00 | 0.61 | 0.58 | 0.29 | 0.49 | 0.32 | 0.40 | 0.32 | 0.11 | 0.03 | 0.51 | 0.51 | 0.49 | 0.51 | 0.49 | 0.51 | 0.49 |
| Ethylbenzene| 1.00 | 0.58 | 0.58 | 0.29 | 0.49 | 0.32 | 0.40 | 0.32 | 0.11 | 0.03 | 0.51 | 0.51 | 0.49 | 0.51 | 0.49 | 0.51 | 0.49 |
| m-, p-Xylenes| 1.00 | 0.58 | 0.58 | 0.29 | 0.49 | 0.32 | 0.40 | 0.32 | 0.11 | 0.03 | 0.51 | 0.51 | 0.49 | 0.51 | 0.49 | 0.51 | 0.49 |
| Styrene| 1.00    | 0.02    | -0.04   | 0.98             | 0.22          | -0.02       | 0.89          | 0.97    | 0.49     | 0.96                   |                 |        |                      |                      |          |                      |
| a-Pinene| 1.00   | 0.32    | 0.05    | 0.85             | 1.00          | 0.05        | 0.51          | 0.97    | 0.49     |                        |                 |        |                      |                      |          |                      |
| 1,3,5-Trimethilbenzene| 1.00 | 0.32 | 0.05 | 0.85 | 1.00 | 0.05 | 0.51 | 0.97 | 0.49 | 0.96 |
| 2-Ethyl-1-hexanol| 1.00 | 0.60 | 0.16 | 0.34 | 0.51 | 0.15 | 0.15 | 0.15 | 0.15 | 0.15 |
| Decane| 1.00    | -0.19   | 0.08    | 0.14          | -0.05        | 0.08        | 0.14          | -0.05   | 0.08     | -0.05                  | -0.05           | 0.08   | -0.05                 | -0.05                 | 0.08     | -0.05                 |
| 1,4-Dichlorobenzene| 1.00 | 0.86 | 0.43 | 0.90 | 1.00 | 0.54 | 0.97 | 1.00 | 0.54 | 0.97 |
| 1,2,4-Trimethylbenzene| 1.00 | 0.54 | 0.97 | 1.00 | 0.54 | 0.97 | 1.00 | 0.54 | 0.97 | 1.00 |
| Limonene| 1.00    | 0.42    |        |                |              |              |              |        |          |                        |                 |        |                      |                      |          |                      |
| 1,2,3-Trimethylbenzene| 1.00 |      |        |                |              |              |              |        |          |                        |                 |        |                      |                      |          |                      |
Table 5: Noncancer reference concentrations, cancer unit risks of the VOCs found during the monitoring campaign and their carcinogenic classifications in the IARC. The mean hazard ratio (HR) and the lifetime cancer risk (LCR) of each compound was calculated.

| VOCs            | Noncancer Reference concentration (µg/m³) | Source | Group IARC | Unit risk (µg/m³)^(-1) | Source | HRs   | LCRs   |
|-----------------|------------------------------------------|--------|------------|------------------------|--------|-------|-------|
| Benzene         | 9.6                                      | ATSDR  | 1          | 6.0E-06                | WHO    | 1.44E-01 | 8.29E-06 |
| Heptane         | —                                        |        |            |                        |        |        |       |
| Toluene         | 5000                                     | IRIS   | 3          |                        |        | 6.10E-04 |       |
| Tetrachloroethylene | 271                                     | ATSDR  | 2A         | 5.9E-06                | OEHHA  | 1.50E-03 | 2.39E-06 |
| Butyl acetate   | —                                        |        |            |                        |        |        |       |
| Ethylbenzene    | 1300                                     | ATSDR  | 2B         | 2.5E-06                | OEHHA  | 5.31E-04 | 1.72E-06 |
| m-, p-Xylenes   | 217                                      | ATSDR  | 3          |                        |        | 9.83E-03 |       |
| o-Xylene        | 217                                      | ATSDR  | 3          |                        |        | 3.82E-03 |       |
| Styrene         | 852                                      | ATSDR  | 2B         |                        |        | 2.48E-04 |       |
| a-Pinene        | —                                        |        |            |                        |        |        |       |
| 1,3,5-Trimethylbenzene | 6                              | PPRTV |            |                        |        | 4.10E-02 |       |
| 2-Ethyl-1-hexanol | —                                        |        |            |                        |        |        |       |
| Decane          | —                                        |        |            |                        |        |        |       |
| 1,4-Dichlorobenzene | 60                                | ATSDR  | 2B         | 1.1E-05                | OEHHA  | 1.09E-03 | 7.21E-07 |
| 1,2,4-Trimethylbenzene | 7                              | PPRTV |            |                        |        | 6.15E-02 |       |
| Limonene        | —                                        |        |            |                        |        |        |       |
| 1,2,3-Trimethylbenzene | 5                              | PPRTV |            |                        |        | 5.68E-02 |       |

IRIS: Integrated Risk Information System; ATSDR: Agency for Toxic Substances and Disease Registry; PPRTV: Provisional Peer Reviewed Toxicity Values of IRIS; OEHHA: Office of Environmental Health Hazard Assessment; WHO: World Health Organization.

Table 6: The lifetime cancer risk (LCR) of each compound and the global lifetime cancer risk in each investigated site.

| Sites | Benzene LCR | Tetrachloroethylene LCR | Ethylbenzene LCR | 1,4-Dichlorobenzene LCR | Global lifetime cancer risk |
|-------|-------------|-------------------------|------------------|-------------------------|-----------------------------|
| Site 1 | 1.10E-05    | 2.26E-06                | 2.37E-06         | 7.66E-07                | 1.64E-05                    |
| Site 2 | 5.99E-06    | 1.77E-06                | 1.06E-06         | 6.38E-07                | 9.45E-06                    |
| Site 3 | 5.29E-06    | 1.59E-06                | 8.17E-07         | 6.96E-07                | 8.40E-06                    |
| Site 4 | 5.36E-06    | 1.68E-06                | 9.50E-07         | 6.88E-07                | 8.69E-06                    |
| Site 5 | 1.54E-05    | 2.09E-06                | 3.53E-06         | 7.94E-07                | 2.18E-05                    |
| Site 6 | 7.29E-06    | 1.69E-06                | 1.42E-06         | 6.99E-07                | 1.11E-05                    |
| Site 7 | 5.35E-06    | 1.71E-06                | 1.09E-06         | 6.90E-07                | 8.84E-06                    |
| Site 8 | 1.28E-05    | 5.20E-06                | 3.10E-06         | 8.22E-07                | 2.20E-05                    |
| Site 9 | 7.71E-06    | 3.45E-06                | 1.71E-06         | 7.21E-07                | 1.36E-05                    |
| Site 10| 8.29E-06    | 2.65E-06                | 1.62E-06         | 7.51E-07                | 1.33E-05                    |
| Site 11| 6.66E-06    | 2.23E-06                | 1.31E-06         | 6.68E-07                | 1.09E-05                    |

to identify the vehicular traffic emissions as the main source of pollution in the urban area of Modugno. Despite that the benzene annual concentration is lower than the limit value, the estimating of the cancer risk showed that the global lifetime cancer risk attributed to investigated VOC exposure was not negligible and therefore should be taken into account in future regulatory approaches. In particular, greater attention should be given to VOCs because they can affect air pollution and the distribution of cancer and non-cancer health risks.
References

[1] E. Vega, G. Sánchez-Reyna, V. Mora-Perdomo et al., “Air quality assessment in a highly industrialized area of Mexico: concentrations and sources of volatile organic compounds,” Fuel, vol. 90, no. 12, pp. 3509–3520, 2011.

[2] P. S. Monks, C. Granier, S. Fuzzi et al., “Atmospheric composition change—global and regional air quality,” Atmospheric Environment, vol. 43, no. 33, pp. 5268–5350, 2009.

[3] E. P. Olauguer, “Near-source air quality impacts of large olefin flares,” Journal of the Air and Waste Management Association, vol. 62, no. 8, pp. 978–988, 2012.

[4] C. Cai, F. Geng, X. Tie, Q. Yu, and J. An, “Characteristics and source apportionment of VOCs measured in Shanghai, China,” Atmospheric Environment, vol. 44, no. 38, pp. 5005–5014, 2010.

[5] R. Atkinson, “Atmospheric chemistry of VOCs and NO(x),” in Reactive Hydrocarbons in the Atmosphere, C. N. Hewitt, Ed., pp. 1–39, Academic Press, San Diego, Calif, USA, 1999.

[6] A. M. Moro, N. Brucker, M. Char´ao et al., “Evaluation of genotoxicity and oxidative damage in painters exposed to low levels of toluene,” Mutation Research, vol. 746, no. 1, pp. 42–48, 2012.

[7] A. Guenther, “A global model of natural volatile organic compound emissions,” Journal of Geophysical Research, vol. 100, no. 5, pp. 8873–8892, 1995.

[8] A. Guenther, T. Karl, P. Harley, C. Wiedinmyer, P. I. Palmer, and C. Geron, “Estimates of global terrestrial isoprene emissions using MEGAN (Model of Emissions of Gases and Aerosols from Nature),” Atmospheric Chemistry and Physics Discussions, vol. 6, no. 1, pp. 107–173, 2006.

[9] K. Na, Y. P. Kim, K. C. Moon, I. Moon, and K. Fung, “Concentrations of volatile organic compounds in an industrial area of Korea,” Atmospheric Environment, vol. 35, no. 15, pp. 2747–2756, 2001.

[10] B. T. Jobson, C. M. Berkowitz, W. C. Kuster et al., “Hydrocarbon source signatures in Houston, Texas: influence of the petrochemical industry,” Journal of Geophysical Research D, vol. 109, no. 24, pp. 1–26, 2004.

[11] C. H. Lai, K. S. Chen, Y. T. Ho, Y. P. Peng, and Y. M. Chou, “Receptor modeling of source contributions to atmospheric hydrocarbons in urban Kaohsiung, Taiwan,” Atmospheric Environment, vol. 39, no. 25, pp. 4543–4559, 2005.

[12] C. Badol, N. Locoge, and J. C. Galloo, “Using a source-receptor approach to characterise VOC behaviour in a French urban area influenced by industrial emissions. Part II: source contribution assessment using the Chemical Mass Balance (CMB) model,” Science of the Total Environment, vol. 389, no. 2–3, pp. 429–440, 2008.

[13] K.-H. Kim, D. X. Ho, C. G. Park et al., “Volatile organic compounds in ambient air at four residential locations in Seoul, Korea,” Environmental Engineering Science, vol. 29, no. 9, pp. 875–889, 2012.

[14] U.S. Department of Health and Human Services, “Toxicological profile for benzene,” Report 200-2004-09793. ATSDR, U.S. Department of Health and Human Services, Atlanta, Ga, USA, 2006.

[15] USEPA and United States Environmental Protection Agency, “EPA air quality trenak,” EPA 454/F-95-003, Office of Air Quality Planning and Standards, USEPA, Research Triangle Park, NC, USA, 1995.

[16] M. A. Mehlm, “Dangerous and cancer-causing properties of products and chemicals in the oil refining and petrochemical industry. VIII. Health effects of motor fuels: carcinogenicity of gasoline-Scientific update,” Environmental Research, vol. 59, no. 1, pp. 238–249, 1992.

[17] H. Fromme, “Significance for the health of the general population of exposure to benzene in traffic,” International Journal of Hygiene and Environmental Health, vol. 196, no. 6, pp. 481–494, 1995.

[18] R. L. Pearson, H. Wachtel, and K. L. Ebi, “Distance-weighted traffic density: in proximity to a home is a risk factor for leukemia and other childhood cancers,” Journal of the Air and Waste Management Association, vol. 50, no. 2, pp. 175–180, 2000.

[19] European Commission, “Directive 2008/50/EC of the European parliament and of the council of 21 May 2008 on ambient air quality and cleaner air for Europe,” Official Journal of the European Union L, vol. 152, pp. 1–44, 2008.

[20] WHO (World Health Organization), Air Quality Guidelines for Europe, WHO Regional Office for Europe, Copenhagen, Denmark, 2nd edition, 2000.

[21] N. Ramírez, A. Cuadras, E. Rovira, F. Borrell, and R. M. Marcé, “Chronic risk assessment of exposure to volatile organic compounds in the atmosphere near the largest Mediterranean industrial site,” Environment International, vol. 39, no. 1, pp. 200–209, 2012.

[22] M. Nadal, M. Schuhmacher, and J. L. Domingo, “Long-term environmental monitoring of persistent organic pollutants and metals in a chemical/petrochemical area: human health risks,” Environmental Pollution, vol. 159, no. 7, pp. 1769–1777, 2011.

[23] K. Rumberger, H. Brown, and J. Spickett, “Volatile organic compounds: do they present a risk to our health?” Reviews on Environmental Health, vol. 22, no. 1, pp. 39–55, 2007.

[24] ECRN (European Chemical Regions Network),” 2010, http://www.ecrn.net/regions/cat.php .

[25] P. Bruno, M. Caputi, M. Caselli, G. De Gennaro, and M. De Rienzo, “Reliability of a BTEX radial diffusive sampler for thermal desorption: field measurements,” Atmospheric Environment, vol. 39, no. 7, pp. 1347–1355, 2005.

[26] M. Caselli, G. De Gennaro, A. Mazzecca, L. Trizio, and M. Tutino, “Assessment of the impact of the vehicular traffic on BTEX concentration in ring roads in urban areas of Bari (Italy),” Chemosphere, vol. 81, no. 3, pp. 306–311, 2010.

[27] P. Bruno, M. Caselli, G. De Gennaro, L. Scolletta, L. Trizio, and M. Tutino, “Assessment of the impact produced by the traffic source on VOC level in the urban area of Canosa di Puglia (Italy),” Water, Air, and Soil Pollution, vol. 193, no. 1–4, pp. 37–50, 2008.

[28] R. R. Hoque, P. S. Khillare, T. Agarwal, V. Shridhar, and S. Balachandran, “Spatial and temporal variation of BTEX in
the urban atmosphere of Delhi, India,” *Science of the Total Environment*, vol. 392, no. 1, pp. 30–40, 2008.

[32] R. Hartmann, U. Voght, G. Baumbach, R. Seyfioglu, and A. Muezzinoglu, “Results of emission and ambient air measurements of VOC in Izmir,” *Environmental Research Funders’ Forum*, vol. 7-8, pp. 107–112, 1997.

[33] A. Muezzinoglu, M. Odabasi, and L. Onat, “Volatile organic compounds in the air of Izmir, Turkey,” *Atmospheric Environment*, vol. 35, no. 4, pp. 753–760, 2001.

[34] A. B. Hansen and F. Palmgren, “VOC air pollutants in Copenhagen,” *Science of the Total Environment*, vol. 189-190, pp. 451–457, 1996.

[35] S. Vardoulakis, N. Gonzalez-Flesca, and B. E. A. Fisher, “Assessment of traffic-related air pollution in two street canyons in Paris: implications for exposure studies,” *Atmospheric Environment*, vol. 36, no. 6, pp. 1025–1039, 2002.

[36] A. Monod, B. C. Sive, P. Avino, T. Chen, D. R. Blake, and F. Sherwood Rowland, “Monoaromatic compounds in ambient air of various cities: a focus on correlations between the xylenes and ethylbenzene,” *Atmospheric Environment*, vol. 35, no. 1, pp. 135–149, 2001.

[37] J. Zhang, T. Wang, W. L. Chameides, C. Cardelino, D. R. Blake, and D. G. Streets, “Source characteristics of volatile organic compounds during high ozone episodes in Hong Kong, Southern China,” *Atmospheric Chemistry and Physics*, vol. 8, no. 16, pp. 4983–4996, 2008.

[38] M. I. Khoder, “Ambient levels of volatile organic compounds in the atmosphere of Greater Cairo,” *Atmospheric Environment*, vol. 41, no. 3, pp. 554–566, 2007.

[39] L. Miller, X. Xu, A. Wheeler, D. O. Atari, A. Grgicak-Mannion, and I. Luginaah, “Spatial variability and application of ratios between BTEX in two Canadian cities,” *The Scientific World Journal*, vol. 11, pp. 2536–2549, 2011.

[40] “IRIS (Integrated Risk Information System),” 2012, http://cfpub.epa.gov/ncea/iris/index.cfm?fuseaction=iris.showsubstanceList.

[41] “ATSDR (Agency for Toxic Substances and Disease Registry),” 2012, http://www.atsdr.cdc.gov/substances/indexAZ.asp#B.

[42] “OEHHAA (Office of Environmental Health Hazard Assessment),” 2003, http://oehha.ca.gov/riskChemicalDB/index.asp.