Local plant adaptation across a subarctic elevational gradient

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1. Summary

Predicting how plants will respond to global warming necessitates understanding of local plant adaptation to temperature. Temperature may exert selective effects on plants directly, and also indirectly through environmental factors that covary with temperature, notably soil properties. However, studies on the interactive effects of temperature and soil properties on plant adaptation are rare, and the role of abiotic versus biotic soil properties in plant adaptation to temperature remains untested. We performed two growth chamber experiments using soils and *Bistorta vivipara* bulbil ecotypes from a subarctic elevational gradient (temperature range: ±3°C) in northern Sweden to disentangle effects of local ecotype, temperature, and biotic and abiotic properties of soil origin on plant growth. We found partial evidence for local adaption to temperature. Although soil origin affected plant growth, we did not find support for local adaptation to either abiotic or biotic soil properties, and there were no interactive effects of soil origin with ecotype or temperature. Our results indicate that ecotypic variation can be an important driver of plant responses to the direct effects of increasing temperature, while responses to covariation in soil properties are of a phenotypic, rather than adaptive, nature.

2. Introduction

Local adaptation of plants to environmental conditions is receiving increasing interest, particularly in the context of climatic change [1,2]. Patterns of local adaptation provide important insights into how plants might respond to shifts in climatic conditions such as global warming. Therefore, many studies have focused on adaptation of plants to temperature. For example, it has been shown that along large-scale latitudinal temperature gradients, plant ecotypes performed best at their home sites [3,4]. However, temperature does not only directly impact on plant performance, and its effects on other ecosystem components may also indirectly affect performance, thereby modifying local plant adaptation. Previous studies have indicated that plant adaptation
to temperature may be coupled with adaptation of aboveground plant-associated organisms to temperature. For example, Laine [5] showed that the relative performance of sympatric versus allopatric populations of fungal pathogens on *Plantago lanceolata* shifted with temperature. Other studies have suggested that plant adaptation to temperature may trade-off against adaptation to insect herbivory. Notably, plants growing at higher elevation would need to allocate fewer resources to herbivore defence because the lower temperatures would moderate herbivore pressure [6].

Few studies, however, have considered interactions with belowground factors [4,7] and we do not know the role of abiotic versus biotic soil properties in determining plant adaptation to temperature. This is despite increased recognition that plant–soil interactions and local adaptation between plants and soil biota can regulate plant growth and fitness [8,9]. Understanding co-adaptation of plants to soil properties and temperature is also important for predicting the potential for evolutionary adaptation to future climatic conditions. Soils generally respond relatively slowly to shifting temperatures [10]; hence, strong local adaptation to soil conditions may impede the ability of plants to track climatic changes (such as those resulting from global warming) which could in turn negatively affect population fitness [11].

We used an elevational gradient in subarctic Sweden to test local plant adaptation to temperature and soil properties. Previous studies in this system have shown that an elevational increase of approximately 500 m corresponds with a decrease in growing season temperature of about 3°C, which makes this system powerful for comparisons with predicted future climate change scenarios [12]. Along this gradient, increasing elevation, and thus declining temperature, causes shifts in abiotic and biotic soil properties [13], which can result from direct effects of temperature or from indirect effects through shifts in plant physiology and community composition [14], with potential consequences for plant performance. In particular, it has been shown that rates of mineralization of soil nutrients, and hence availability of nutrients for plant uptake, generally decline with elevation [13,15]. To further our understanding of plant adaptive responses to temperature, it is therefore imperative to consider interactions with both abiotic and biotic soil properties.

We tested patterns of local adaptation for *Bistorta vivipara*, a viviparous, perennial, bulbil-producing plant species commonly found in (sub)arctic and alpine habitats [16]. Soils and *B. vivipara* bulbils were collected from three elevations (i.e. three temperature habitats) and seedling responses were tested in two fully reciprocal experiments (with ecotype, temperature and soil origin as factors), using climate-controlled growth chambers to mimic conditions of each temperature habitat. We tested the following hypothesis: local adaptation of *B. vivipara* to temperature is coupled with adaptation to associated soil properties, due in part to co-adaptation of plants and soil organisms. To diagnose local adaptation, we used two non-independent criteria [1,2]: (i) the ‘local versus foreign contrast’ (comparing performance of focal versus other ecotypes in a focal habitat), and (ii) the ‘sympatric versus allopatric contrast’ (comparing performance of a focal ecotype in focal versus other habitats).

### 3. Material and methods

#### 3.1. Study site

Soils and *B. vivipara* (syn: *Polygonum viviparum*) bulb ecotypes used in this study were collected from an elevational gradient on the northeast facing slope of Mount Suorooaivi (1193 m), approximately 20 km southeast of Abisko, Sweden (68°21′ N, 18°49′ E) [13,15]. The climate is subarctic with long, cold winters and a short growing season from June to September. The average maximum temperature at 450, 700 and 900 m during the 2012 growing season was 13.5°C, 12.3°C and 9.7°C, respectively; the average minimum temperature corresponded to 5.0°C, 4.1°C and 3.8°C (figure 1a). Mean annual precipitation is approximately 310 mm (1913–2000, Abisko Scientific Research Station). Parent soil material is composed of salic igneous rocks and quartic and phyllitic hard schists. The vegetation typically consists of mosaics of two co-dominant types: meadow and heath, which occur across the gradient. Meadow vegetation is diverse and dominated by forbs, graminoids and sedges, while heath vegetation is less diverse and dominated by deciduous and evergreen dwarf-shrubs and *Betula nana*. *Bistorta vivipara* typically occurs in meadow vegetation [13].

#### 3.2. Soil collection

Soils for our experiments were collected in August 2012 at 450, 700 and 900 m elevations. Soils were collected to 10 cm depth adjacent to five existing 1 × 1 m meadow plots at each elevation established
Figure 1. (a) Minimum and maximum daily air temperature from early July to mid-September 2012 at 450, 700 and 900 m along the elevational gradient used for our study in Abisko, Sweden. Closed points represent daily maximum temperature. Open points represent daily minimum temperature. Closed and dotted lines show trend lines from quadratic regressions for daily maximum and minimum temperatures, respectively. (b) Temperature settings for climate chambers used for testing seedling responses mimicking the temperature conditions at 450, 700 and 900 m as observed along the elevational gradient. Weekly average maximum (closed lines) and minimum temperature values (dotted lines) were derived from equations of the trend lines of regression as shown in (a) with the exception that the minimum temperature could not be lower than 5°C because of constraints of the growth chambers used (grey dotted line; see Material and methods). Diurnal temperature cycles were set by hourly increments between minimum and maximum values (not depicted in the figure).

earlier in 2012. The mean distance between plots within elevations was ca 48 m with the maximum distance between the two plots furthest apart being ca 140 m. Because of high spatial heterogeneity over short distances in micro-topography, hydrology and soil fertility in these plant communities [17], it is expected that this distance among plots is sufficient to ensure adequate independence among them [18]. After collection, the soils were frozen (−18°C) for six weeks until they were further processed for our experiments; these freezing conditions are well within the range of what these soils experience during the winter.
3.3. Soil abiotic and biotic properties

In July 2012, we collected three to six soil cores from each of the five 1 x 1 m meadow plots at each elevation (45 mm diameter, 10 cm depth) for analysis of abiotic and biotic soil properties. Cores were bulked within plots and kept at 4°C overnight before being passed through a 4 mm mesh sieve to remove plant matter and stones. Soil pH was determined on a subsample of fresh soil (2.5 g dry weight) after shaking for 12 h in 40 ml deionized water. Gravimetric moisture content was determined after drying (105°C, 24 h) and soil organic matter content was determined after combustion in a muffle furnace (550°C, 4 h). A subsample of soil was dried (60°C, 72 h), ground and analysed for total carbon (C) and nitrogen (N) by dry combustion using a FLASH 2000 Organic Elemental Analyzer (Interscience, Breda, The Netherlands) and phosphorus (P) with nitric-perchloric acid digestion analysed by inductively coupled plasma [19]. For each subsample of soil, microbial communities were characterized by phospholipid fatty acid (PLFA) analysis [20,21] with different PLFAs representative of different groups of soil microorganisms. We extracted PLFAs from each subsample after freeze-drying and grinding, according to Frostegård et al. [22]. We used i14 : 0, 14 : 0, i15 : 0, a15 : 0, 15 : 0, i16 : 0, 16 : 1ω9, 16 : 1ω7c, 16 : 1ω7t, i17 : 0, a17 : 0, 17 : 1ω8, cy17 : 0, 17 : 0, 18 : 1ω7, cy19 : 0 as indicators for bacteria and 18 : 2ω6 as an indicator for fungi.

3.4. Bistorta ecotypes

*Bistorta vivipara* is a viviparous perennial knotweed commonly found in (sub)arctic and alpine habitats [16,23–25]. Vivipary has been suggested as an adaptation to cold climates [26,27], and in the subarctic tundra of northern Sweden viviparous inflorescences are very common, while the number of flowers is generally low. Occasional sexual reproduction does occur [16,28], but *B. vivipara* normally reproduces asexually by bulbils (vegetative reproduction organs), and this should serve to limit gene flow. Notably, the potential for local adaptation is highest for ‘intermediate’ levels of gene flow [29,30]. *Bistorta vivipara* clones can be widespread [25], but several studies have shown that genotypic variation within and across *B. vivipara* populations in arctic and alpine zones is rather high, with many clones being unique to single populations [16,31,32]. As such, we considered *B. vivipara* a good candidate species to address our hypothesis. At each elevation, *B. vivipara* bulbils were collected within 25 m of the plots. Bulbils were collected between mid-August and early September 2012 from approximately 200 individuals per elevation. These bulbils were stored at 4°C for a maximum of six weeks, after which they were surface-sterilized (1% sodium hypo-chloride solution for 1 min) [33], and germinated on sterile soil under common conditions in a greenhouse for use in our growth chamber experiments. Additionally, for each elevation, 50 randomly selected bulbils were oven-dried and weighed in order to compare dry weight among ecotypes.

3.5. Experimental design

The soils and bulbil ecotypes were then used in two growth chamber experiments. Experiment 1 addressed whether local adaptation of *B. vivipara* to temperature is coupled with adaptation to associated soil properties. This experiment included three fully crossed factors: soil origin (from 450, 700 and 900 m elevations), bulbil ecotype (from 450, 700 and 900 m elevations) and temperature (mimicking regimes at 450, 700 and 900 m elevations), with five replicates of each treatment combination (3 x 3 x 3 x 5 replicates = 135 experimental units). Soils were processed by removing large stones and roots and then homogenized per elevation, because we were interested in local adaptation to soil properties related to elevation, rather than small-scale within-site heterogeneity [34]. For each treatment combination, plastic pots (8 x 8 x 10 cm deep, each containing a drainage layer of 150 ml sterilized sand) were filled with the field soil and planted with two pre-germinated *B. vivipara* seedlings.

For each soil origin x ecotype combination, five replicate pots were placed in each of three growth chambers, whose conditions were set to mimic temperature conditions as observed at 450, 700 and 900 m along the elevational gradient during the 2012 growing season (figure 1). Average minimum and maximum temperatures over this period were 5.0 and 13.5°C for the 450 m elevation, 4.1 and 12.3°C for the 700 m elevation, and 3.8 and 9.7°C for the 900 m elevation. Weekly average maximum and minimum temperatures were derived from equations of the polynomial trend lines of regressions as shown in figure 1a. Diurnal temperature cycles in the growth chambers were set by hourly increments between minimum and maximum values. Minimum temperature values lower than 5°C were set to 5°C owing to climate chamber system constraints. Relative humidity was set to 90% in all chambers, and
photosynthetically active radiation (PAR) and hours of daylight were set in each growth chamber to represent conditions measured from mid-June to mid-September 2012 in the vicinity of the elevational gradient used in our study (using data derived from the Abisko Scientific Research Station for PAR and from the Swedish Meteorological and Hydrological Institute for day length; electronic supplementary material, figure S1). Each of the five replicates of each treatment combination was randomly assigned to a position in a block within each chamber. All pots were isolated in individual saucers to prevent cross-contamination through water movement. Pots were rotated systematically between chambers (i.e. entire temperature treatments moved between chambers) on a weekly basis to minimize effects of any uncontrolled sources of variation among chambers. Pots were watered as necessary throughout the experiment to eliminate moisture as a limiting factor.

Experiment 2 addressed whether local adaptation to temperature is due in part to co-adaptation of plants and soil organisms; it differentiated between the effects of abiotic and biotic soil properties using sterilized soils and living soil inocula. By comparing sterilized soils to re-inoculated soils, the effects of abiotic versus biotic soil factors on plant growth can be effectively separated [33,35]. The experiment followed a full-factorial design including inoculum origin (inoculum from 450, 700 and 900 m elevations, and a non-inoculated sterile control), bulbil ecotype (from 450, 700 and 900 m elevations) and temperature (mimicking regimes at 450, 700 and 900 m elevations), with five replicates of each treatment combination (4 × 3 × 3 × 5 replicates = 180 experimental units). Soils were processed as described for experiment 1. We then took an equal subsample of soil from each of the three elevations and combined them into one mixed bulk soil which we sent to Synergy Health (Swindon, UK) for γ-irradiation (25 kGy), which is widely used for killing soil biota [33]. We prepared four inoculation treatments, i.e. a ‘control’ sterilized soil without any inoculation, or inoculation of the sterilized soil with live soil from each of the three elevations. For each of the three treatments in which inoculation was performed, a volume of 10% live soil from one of the three elevations was mixed with the bulked sterilized soil; for the ‘control’ soil only bulked sterilized soil was used. The soil inocula potentially included all soil microflora (prokaryotes and fungi), and micro- and mesofauna. The inoculated soils were then placed in pots and planted with B. vivipara seedlings as described for experiment 1. The experiment was then run following approaches and methodology described for experiment 1.

For both experiments, plants were grown for 12 weeks, following temperature and light conditions characteristic for the Abisko growing season (figure 1; electronic supplementary material, figure S1). Experiments like this are usually run for the duration of several months [33,36], and typically focus on the seedling stage because seedling performance is an important component of plant fitness [37]. At harvest, soil was carefully washed from the roots and roots and shoots were separated. Shoot and root dry mass were determined after drying at 60°C for at least 72 h.

3.6. Data analyses

The effects of ecotype, temperature and soil/inoculum origin, and their interactions, on total plant biomass were tested using three-way ANOVA. Biomass of the two seedlings per pot was averaged, and pots served as the unit of replication (n = 5). To assess significant differences among means, Tukey’s HSD post hoc tests with Bonferroni corrections (p = 0.05) were run for (i) ‘local versus foreign’ comparisons, and (ii) ‘sympatric versus allopatric’ comparisons [1], i.e. pairwise comparison of the levels of each factor within the levels of the other factor in the interaction [38]. Effects of ecotype on bulbil dry weight and effects of elevation on abiotic and biotic soil properties were tested using one-way ANOVA with Tukey’s HSD post hoc tests. For all data analyses, data were transformed when necessary to meet the assumptions for parametric testing. All statistical analyses were performed in SPSS (PASW statistics v. 21.0, IBM Corporation, Armonk, NY, USA).

4. Results

4.1. Experiment 1

There was a significant interactive effect between ecotype and temperature on plant biomass, but there were no interactions of soil origin with ecotype or temperature (table 1). Furthermore, plant biomass was higher in the 450 m soil than the 700 and 900 m soils (figure 2b). Local versus foreign contrast (figure 2b): at the 450 m temperature, biomass of the 450 m ecotype was higher than for the 700 and 900 m ecotypes; at the 700 m temperature, biomass of the 450 and 700 m ecotypes was higher than the 900 m ecotype; and at the 900 m temperature, biomass of the 450 m ecotype was higher than the 900 m ecotype. Sympatric
experiment 1: living soils

(a) Biomass responses in living soils under temperature regimes associated with 450, 700 and 900 m elevations. Different capital letters denote significant differences among ecotypes (a,c) or soil origins (b). Different lower case letters denote significant differences among temperature treatments within ecotypes.

(b) Biomass responses in living soils collected from 450, 700 and 900 m elevations, averaged across all temperatures. Different capital letters denote significant differences (Tukey’s HSD post hoc tests, p < 0.05) among ecotypes (a,c) or soil origins (b). Different lower case letters denote significant differences among temperature treatments within ecotypes.

experiment 2: inoculated soils

(c) Biomass responses in sterilized soils re-inoculated with living soil inoculum under temperature regimes associated with 450, 700 and 900 m elevations. (d) Biomass responses in sterilized soils re-inoculated with living soil inoculum collected from 450, 700 and 900 m elevations, and in sterilized soil without inoculum (control), averaged across all temperatures. Different capital letters denote significant differences among ecotypes (a,c) or soil origins (b). Different lower case letters denote significant differences among temperature treatments within ecotypes.

Figure 2. Biomass responses (mean ± s.e., n = 5) of Bistorta vivipara ecotypes under temperature and soil treatments (local versus foreign contrast). Bulbil ecotypes were collected from 450 m (E450), 700 m (E700) and 900 m (E900) elevations and grown under temperature regimes (‘temp’) associated with 450, 700 and 900 m elevations. (a) Biomass responses in living soils under temperature regimes associated with 450, 700 and 900 m elevations. (b) Biomass responses in living soils collected from 450, 700 and 900 m elevations, averaged across all temperatures. (c) Biomass responses in sterilized soils re-inoculated with living soil inoculum under temperature regimes associated with 450, 700 and 900 m elevations. (d) Biomass responses in sterilized soils re-inoculated with living soil inoculum collected from 450, 700 and 900 m elevations, and in sterilized soil without inoculum (control), averaged across all temperatures. Different capital letters denote significant differences (Tukey’s HSD post hoc tests, p < 0.05) among ecotypes (a,c) or soil origins (b). Different lower case letters denote significant differences among temperature treatments within ecotypes.

versus allopatric contrast (electronic supplementary material, figure S2a): for the 450 and 700 m ecotypes, biomass was highest at temperature treatments corresponding to their elevation of origin; and there was no temperature effect on the ecotype from 900 m.

4.2. Experiment 2

Plant biomass patterns in experiment 2 were generally similar to those observed in experiment 1. Local versus foreign contrast (figure 2c): at the 450 m temperature, the 450 m ecotype had significantly higher biomass than did the 700 and 900 m ecotypes; similarly, at the 700 m temperature, the 700 m ecotype had a significantly higher biomass than did the two other ecotypes; and at the 900 m temperature there were no significant biomass differences between the three ecotypes (figure 2b). Plant biomass was not affected by inoculum origin, or by its interaction with ecotype and/or temperature (figure 2d). Sympatric versus allopatric contrast (electronic supplementary material, figure S2b): for the 450 and 700 m ecotypes, biomass was highest at temperature treatments corresponding to their elevation of origin; and there was no temperature effect on the ecotype from 900 m. Biomass of the 450 and 700 m ecotypes differed significantly from that of the 900 m ecotype, but not from each other.
Table 1. Results from ANOVA of effects of ecotype (450, 700, 900 m), temperature regime (450, 700, 900 m) and soil or inoculum origin (450, 700, 900 m; control (experiment 2 only)) on total biomass of *Bistorta vivipara* grown on living and inoculated soils. (Significant *p*-values (*<* 0.05) are in bold type. Non-significant two- and three-way interactions are not shown.)

| effect          | living soils |          |          | inoculated soils |          |          |
|-----------------|--------------|----------|----------|-----------------|----------|----------|
|                 | DF<sub>num</sub> | *F*-value | *p*-value | DF<sub>num</sub> | *F*-value | *p*-value |
| ecotype (*E*)   | 2            | 31.78    | <0.001   | 2               | 20.64    | <0.001   |
| temperature (*T*)| 2            | 25.75    | <0.001   | 2               | 21.55    | <0.001   |
| soil origin (*S*)| 2           | 21.07    | <0.001   |                 |          |          |
| inoculum origin (*I*) | 3        | 0.50     | 0.69     |                 |          |          |
| *E* × *T*      | 4            | 9.20     | <0.001   | 6               | 11.22    | <0.001   |
| error          | 108          |          |          | 144             |          |          |

Figure 3. Biomass (mean ± s.e., *n* = 50) of bulbils of *Bistorta vivipara* ecotypes collected from 450, 700 and 900 m elevations. Bulbil biomass was significantly affected by elevation (ANOVA: *F*<sub>2,147</sub> = 86.47, *p* < 0.001). Different capital letters denote significant differences (Tukey’s HSD post hoc tests, *p* < 0.05) among elevations.

5. Discussion

Disentangling the importance of covarying factors enhances our understanding of plant adaptation to local environmental conditions. We found partial evidence for local adaptation of *B. vivipara* to shifting conditions with elevation, largely manifested by adaptation to temperature. Ecotypes from the lowest elevation performed best at local temperatures (i.e. sympatric versus allopatric contrast), and at temperatures representing low elevation the local ecotype performed best (i.e. local versus foreign contrast). Similar patterns were found for the mid-elevation ecotype and temperatures. These patterns could not solely be explained by the more favourable temperature conditions at lower elevations (i.e. intrinsic differences in habitat quality [2]). Notably, mid-elevation ecotypes did better at mid-elevation temperature which is on average 1.4°C lower than low elevation temperature; this is despite observations that growth of *B. vivipara* in the subarctic is temperature limited [23,24,39]. Moreover, while maternal effects could have contributed to enhanced performance of local ecotypes at low elevation (i.e. bulbils from 450 m were larger than bulbils from other elevations; figure 3), these effects could not account for the ‘local versus foreign’ comparison at mid elevation. In this light, ecotypes from mid elevation performed better at mid-elevation temperatures than at more favourable low-elevation temperatures. Taken together, although we do not know the level of clonal diversity between bulbil ecotypes, the large differences among ecotypes in biomass response to temperature strongly suggest a genetic basis for the observed patterns.

Contrary to our hypothesis, we found that plant adaptation to temperature was independent of soil conditions. Overall, plants performed best in low elevation soils, probably because of increased
soil fertility at lower elevations [13,40]. However, contrary to previous studies [9,41,42], we found no evidence of local adaptation to soil abiotic and/or biotic properties. It has been suggested that local adaptation to soil conditions might be particularly important in ‘extreme’ soil types [4], as has been shown, for example, in nutrient-poor serpentine soils [43]. As such, low soil fertility at the highest elevation [13] would be expected to exert a strong selection pressure. However, in subarctic ecosystems, spatio-temporal variation in soil nutrient availability is strong even within elevations [17], which could work against local adaptation to soil conditions [4]. Moreover, plants were unresponsive to soil biota regardless of soil origin. Strong environmental control (i.e. temperature, soil abiotic properties) of plant growth may have suppressed effects of soil biota, or effects of mutualists and pathogens may have cancelled each other out.

Understanding local adaptation of plants to environmental conditions is important in predicting plant performance and population dynamics under future climate change scenarios [44]. Our results show that with shifts in elevation, plants adapt to local environmental conditions, but that adaptation was driven solely by responses to variation in temperature. Plant growth was responsive to variation in associated soil properties, but these responses were decoupled from local adaptation to temperature. As such, our results indicate that the direct effects of global warming on plant performance depend on the genetic variability in temperature optimum, and would not be impeded by soil legacies. As such, our study sheds new light on the relative importance of genetic adaptation and phenotypic plasticity for plant performance under shifting climatic conditions.

Data accessibility. The datasets supporting this article have been deposited in the Dryad repository [45]: doi:10.5061/dryad.v2k50.

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References

1. Blanquart F, Kaltz O, Huimer SL, Gandon S. 2013 A practical guide to measuring local adaptation. *Ecol. Lett.* 16, 1195–1205. (doi:10.1111/ele.12150)

2. Kawecki TJ, Ebert D. 2004 Conceptual issues in local adaptation. *Ecol. Lett.* 7, 1225–1241. (doi:10.1111/j.1461-0248.2004.00684.x)

3. De Frenne P et al. 2011 Temperature effects on forest herbs assessed by warming and transplant experiments along a latitudinal gradient. * Glob. Change Biol.* 17, 3240–3253. (doi:10.1111/j.1365-2486.2011.02449.x)

4. Macel M et al. 2007 Climate vs. soil factors in local adaptation of two common plant species. *Ecology* 88, 424–433. (doi:10.1890/0012-9658(2007)88[424:cv]2.0.co;2)

5. Laine AL. 2008 Temperature-mediated patterns of local adaptation in a natural plant-pathogen metapopulation. *Ecol. Lett.* 11, 327–337. (doi:10.1111/j.1461-0248.2007.00146.x)

6. Raffa KE, Powell EN, Townend PA. 2013 Temperature-driven range expansion of an invasive insect heightened by weakly coevolved plant defenses. *Proc. Natl Acad. Sci. USA* 110, 2193–2198. (doi:10.1073/pnas.1216668110)

7. De Long JR, Kardol P, Sundqvist MK, Veen GE, Wardle DA. In press. Plant growth response to direct and indirect temperature effects varies by vegetation type and elevation in subarctic tundra. *Oikos* (doi:10.1111/oik.00764)

8. Goh CH, Vallejoa DVF, Nicothal AB, Mathesius U. 2013 The impact of beneficial plant-associated microbes on plant phenotypic plasticity. *J. Chem. Ecol.* 39, 826–839. (doi:10.1007/s10886-013-0326-8)

9. Lau JA, Lennon JT. 2012 Rapid responses of soil microorganisms improve plant fitness in novel environments. *Proc. Natl Acad. Sci. USA* 109, 14 058–14 062. (doi:10.1073/pnas.1203191109)

10. Sistla SA, Rastetter EB, Schimel JP. 2014 Responses of a tundra system to warming using SCAMPs: a stoichiometrically coupled, acclimating microbe–plant–soil model. *Ecol. Monogr.* 84, 151–170. (doi:10.1890/12-1191.1)

11. Schiffer K, Bourne EC, Lavergne S, Thuiller W, Travis JM. 2013 Limited evolutionary rescue of locally adapted populations facing climate change. *Phil. Trans. R. Soc. B* 368, 2020083. (doi:10.1098/rstb.2012.0063)

12. IPCC. 2013 Climate change 2013: the physical science basis. In Contribution of Working Group I to the Fifth Assessment Report of the Intergovernmental Panel on Climate Change (eds TF Stocker et al.), pp. 1–30. Cambridge, UK: Cambridge University Press.

13. Sundqvist MK, Giesler R, Graae BJ, Wallander H, Fogelberg E, Wardle DA. 2011 Interactive effects of vegetation type and elevation on aboveground and belowground properties in a subarctic tundra. *Oikos* 120, 128–142. (doi:10.1111/j.1600-0706.2010.18811.x)

14. Kardol P, Greger MA, Campany CE, Classen AT. 2010 Soil ecosystem functioning under climate change: plant species and community effects. *Ecology* 91, 767–781. (doi:10.1890/09-0135.1)

15. Sundqvist MK, Liu Z, Giesler R, Wardle DA. 2014 Plant and microbial responses to nitrogen and phosphorus addition across an elevational gradient in subarctic tundra. *Ecology* 95, 1819–1835. (doi:10.1890/13-0869.1)

16. Baurt MR. 1996 Genetic diversity and ecotypic differentiation in arctic and alpine populations of Polygonum viviparum. *Arct. Alp. Res.* 28, 190–195. (doi:10.2307/1551759)

17. Bjork RG, Klemetsdottir L, Molau U, Hamborg JF, Odman A, Giesler R. 2007 Linkages between N turnover and plant community structure in a tundra landscape. *Plant Soil* 294, 247–261. (doi:10.1007/s11104-007-0250-4)

18. Sundqvist MK, Wardle DA, Olefson E, Giesler R, Gundale MJ. 2012 Chemical properties of plant litter in response to elevation: subarctic vegetation challenges phenolic allocation theories. *Funct. Ecol.* 26, 1090–1099. (doi:10.1111/ffe.12083)

19. Sparks DL. 1996 Methods of soil analysis, part 3, chemical methods: SSA S5 Book, Series No. 5. Madison, WI: SSA.

20. Bligh EG, Dyer WJ. 1959 A rapid method of total lipid extraction and purification. *Can. J. Biochem. Physiol.* 37, 911–917. (doi:10.1139/o-59-099)

21. White DC, Davis WM, Nickels JS, King JD, Bobbie RJ. 1979 Determination of the sedimentary microbial biomass by extractable lipid phosphate. *Oecologia* 40, 51–62. (doi:10.1007/bf00388380)

22. Frostegård A, Bååth E. 1996 The use of phospholipid fatty acid analysis to estimate bacterial and fungal biomass in soil. *Biol. Fertil. Soils* 22, 59–65. (doi:10.1007/bf00384433)

23. Gugerli F, Baurt MR. 2001 Growth and reproduction of Polygonum viviparum show weak responses to experimentally increased temperature at a Swiss alpine site. *Bot. Helv.* 111, 169–180.
24. Wookey PA, Welker JM, Parsons AN, Press MC, Callaghan TV, Lee JA. 1994 Differential growth, allocation and photosynthetic responses of Polygonum viviparum to simulated environmental change at a high arctic polar semi-desert. Oikos 70, 131–139. (doi:10.2307/3545708)

25. Marr KL, Allen GA, Hebda RJ, McCormick LJ. 2013 Phylogeographical patterns in the widespread arctic-alpine plant Bistorta vivipara (Polygonaceae) with emphasis on western North America. J. Biogeogr. 40, 847–856. (doi:10.1111/jbi.12042)

26. Molau U. 1993 Relationships between flowering phenology and life-history strategies in tundra plants. Arct. Antarct. Alp. Res. 25, 391–402. (doi:10.2307/1551922)

27. Steiner BL, Armbruster GFJ, Scheepens JF, Stocklin J. 2012 Distribution of bulbil- and seed-producing plants of Poa alpina (Poaceae) and their growth and reproduction in common gardens suggest adaptation to different elevations. Am. J. Bot. 99, 2035–2044. (doi:10.3732/ajb.1200213)

28. Söyrinki N. 1989 Fruit production and seedlings in Polygonum viviparum. Memo. Soc. Fauna Flora Fenn. 65, 13–15.

29. Blanquart F, Gandon S. 2011 Evolution of migration in a periodically changing environment. Am. Nat. 177, 188–201. (doi:10.1086/657953)

30. Gandon S. 2002 Local adaptation and the geometry of host-parasite coevolution. Ecol. Lett. 5, 246–256. (doi:10.1046/j.1461-0248.2002.00305.x)

31. Diggle PK, Lower S, Ranker TA. 1998 Clonal diversity in alpine populations of Polygonum viviparum (Polygonaceae). Int. J. Plant Sci. 159, 606–615. (doi:10.1086/297579)

32. Lundqvest E, Andersson E. 2001 Genetic diversity in populations of plants with different breeding and dispersal strategies in a free-flowing boreal river system. Hereditas 135, 75–83. (doi:10.1111/j.1600-0522.2001.00075.x)

33. Kardol P, Cornelis NJ, Van Kempen MML, Bakx-Schotman JMI, Van der Putten WH. 2007 Microbe-mediated plant–soil feedback causes historical contingency effects in plant community assembly. Ecol. Manage. 27, 47–162. (doi:10.1890/06.0502)

34. Pizarro C, Mangan SA, Herre EA, Eom AH, Dalling JW. 2011 Above- and belowground interactions drive habitat segregation between two cryptic species of tropical trees. Ecology 92, 47–56. (doi:10.1890/09-1715.1)

35. Gundale MJ, Kardol P, Nilsson M-C, Nilsson U, Lukas RW, Wardle DA. 2014 Interactions with soil biota shift from negative to positive when a tree species is moved outside its native range. New Phytol. 202, 415–421. (doi:10.1111/nph.12699)

36. Packer A, Clay K. 2000 Soil pathogens and spatial patterns of seedling mortality in a temperate tree. Nature 404, 278–281. (doi:10.1038/35005072)

37. Agrawal AA, Johnson MTJ, Hastings AP, Maron JL. 2013 A field experiment demonstrating plant life-history evolution and its eco-evolutionary feedback to seed predator populations. Am. Nat. 181, 535–545. (doi:10.1086/666727)

38. Meyers LS, Garrot G, Guarion A. 2006 Applied multivariate research: design and interpretation. Thousand Oaks, CA: Sage Publishers.

39. Totland Ø, Hjeltn J. 1998 Assessment of the effects of environmental change on the performance and density of Bistorta vivipara: the use of multivariate analysis and experimental manipulation. J. Ecol. 86, 969–998. (doi:10.1046/j.1365-2745.1998.00318.x)

40. Svenningsson B, Davis J, Abadie W, Butler A. 1995 Soil carbon and nitrogen mineralization at different elevations in the Chugach Mountains of south-central Alaska, USA. Arct. Alp. Res. 27, 29–37. (doi:10.2307/1552865)

41. Lankau RA. 2013 Species invasion alters local adaptation to soil communities in a native plant. Ecology 94, 32–40. (doi:10.1890/12-0673.1)

42. Smith DS, Schweitzer JA, Turk P, Bailey JK, Hart SC, Shuster SM, Whitham TG. 2012 Soil-mediated local adaptation alters seedling survival and performance. Plant Soil 352, 243–251. (doi:10.1007/s11104-011-1092-7)

43. Wright JW, Stanton ML, Scherson R. 2006 Local adaptation to serpentine and non-serpentine soils in Calochortus parviflorus. Evol. Ecol. Res. 8, 1–21.

44. Berg MP, Kiers ET, Driessen G, van der Heijden M, Kooi BW, Kuemen F, Liefting M, Verhoef HA, Eilers J. 2010 Adapt or disperse: understanding species persistence in a changing world. Glob. Change Biol. 16, 587–596. (doi:10.1111/j.1365-2486.2009.02049.x)

45. Data from: local plant adaptation across a subarctic elevational gradient. Data identifier: doi:10.5061/dryad.v2k50 Journal manuscript number: rsos-140141