Spatially-explicit valuation of coastal wetlands for cyclone mitigation in Australia and China

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Coastal wetlands are increasingly recognised for their pivotal role in mitigating the growing threats from cyclones (including hurricanes) in a changing climate. There is, however, insufficient information about the economic value of coastal wetlands for cyclone mitigation, particularly at regional scales. Analysis of data from 1990–2012 shows that the variation of cyclone frequencies is related to El Niño strength in the Pacific Ocean adjacent to Australia, but not China. Among the cyclones hitting the two countries, there are significant relationships between the ratio of total economic damage to gross domestic production (TD/GDP) and wetland area within cyclone swaths in Australia, and wetland area plus minimum cyclone pressure despite a weak relationship in China. The TD/GDP ratio is significantly higher in China than in Australia. Despite their extensive and growing occurrence, seawalls in China appear not to play a critical role in cyclone mitigation, and cannot replace coastal wetlands, which provide other efficient ecosystem services. The economic values of coastal wetlands in Australia and China are respectively estimated at US$52.88 billion and 198.67 billion yr⁻¹ for cyclone mitigation, albeit with large within-country geographic variation. This study highlights the urgency to integrate this value into existing valuations of coastal wetlands.

Coastal wetlands play a pivotal environmental and economic role in protecting coastlines from cyclones, mitigating climate change through carbon sequestration and storage, serving a nursery function for fisheries or recreation function for humans, and mediating the effects of floods or enhancing water quality1–5. Protection from cyclones is paramount due to the overwhelming trend of coastal urbanisation6,7. Coastal wetlands attenuate the potential catastrophic damage of cyclones, relying on their ability to dissipate storm energy8–10. The natural shield provided by wetlands outweighs engineering measures, such as seawalls and bulkheads11,12, which have a defined lifespan, require continual maintenance, and could not replace the fishing, recreational and other roles of coastal wetlands. Such structures can also be costly to build, and exacerbate coastal erosion or cause knock-on effects in neighbouring regions13,14. Despite these shortcomings, hard-engineering options such as seawalls are often the preferred options, resulting in extensive irreversible modification of the coastline, e.g. the ‘New Great Wall’ of China15. Nonetheless, so far, no studies have quantitatively determined the advantage of coastal wetlands over seawalls from the perspective of ecosystem services.

Valuations of wetland protection from coastal natural disasters (including cyclones) generally fall into two classes: ecological or functional valuation and economic or monetary valuation. Increasing academic attention has been paid to the role of wetlands in shielding coastal regions from cyclones since the 2004 Indian Ocean tsunami. Studies have assessed the impact of coastal natural disasters based on anecdotal evidence16, post-hoc observations17, remote sensing, and modelling18. The validity of such studies is limited by the ability to constrain confounding factors. Thus, there is on-going debate over the conclusions of such studies. Lee, et al.19 stated that remote sensing may be ineffective in detecting cryptic degradation, resulting in fragmented mangroves, which may not provide the protection that intact mangroves can. Pinsky, et al.19 stressed the controversial function of coastal wetlands for storm protection due to the variation of geomorphic, ecological and hydrodynamic factors that determine wave attenuation among locations and times. Further, most studies used whole countries as units...
of observation, and may thus miss important spatial variations within a country. It is paramount to conduct studies focusing on a spatially explicit evaluation of the role of wetlands in protection from cyclones at both between-country and within-country scales.

On the Western Pacific rim, China and Australia are the largest countries in the northern and southern hemispheres, respectively; both with long coastlines. Coastal regions of both countries are subject to intense cyclone activities with potentially devastating ecological and economic effects. However, these two countries differ in their natural assets potentially protecting them from cyclones, with large differences in geographic extent and distribution of natural wetlands, ocean meteorological conditions and mitigation options. For example, China has established seawalls along the coastline, reaching 11,000 km in extent by 2010, as an alternative to natural wetlands for protecting the coastal regions from cyclones. Such infrastructure is uncommon in Australia, except some urban areas. On the other hand, Australia regularly suffers from the significant direct impact of El Niño Southern Oscillation, e.g. sustained drought. Further, based on the recent global maps of mangroves and saltmarshes, in coastal wetlands of China, the distribution of mangroves is constrained to subtropical regions, while the distribution of saltmarshes covers almost the whole coastline. In contrast, mangroves and saltmarshes occur almost concomitantly along the coastline in Australia.

Costanza, et al. pioneered the study of valuing wetlands in mitigating hurricanes in the USA over the period 1980–2004. However, their study was restricted to the land use map of one year (i.e. 2000) to extract decadal wetland area. Their study also used only wind speed as an indicator of cyclone intensity, while barometric pressure predicts damage more accurately than maximum sustainable wind speed. Barbier recommended the method of estimating protective services of wetlands in terms of the reduction in the expected damages or deaths inflicted on coastal communities.

The objectives of this study are to: (a) analyse different characteristics (including frequencies and minimum central pressure) of cyclones occurring in oceans around Australia and China during 1990–2012; (b) analyses of the relationship between TD/GDP and wetland area, wind speed, and the duration and minimum central pressure of cyclones for both countries separately; (c) assess the value of Australian and Chinese wetlands in protecting the countries from cyclones over the period 1990–2012; (d) assess the influence of El Niño on cyclone activities in both countries; (e) the effect of seawalls on TD from cyclones in China, in combination with a cost-benefit analysis to compare seawalls and coastal wetlands in their efficacy for cyclone mitigation. In this study, we followed the avenue of Costanza, et al. who estimated the value of wetlands in mitigating hurricanes hitting the USA from the relationship between the ratio of total economic damage to gross domestic production (TD/GDP) and influential variables. However, we refined their method by including the minimum central pressure as an indicator of cyclone intensity (for both China and Australia in the initial regression model) rather than maximum wind speed alone. We used time-series land-use maps in our analysis and also assessed the effect of El Niño on cyclone occurrence, as well as the impact of seawalls on cyclone mitigation in China.

**Results**

We assessed the variation of cyclone frequencies and intensities along with the occurrence of El Niño events. Of the 531 cyclones in Australia and China during 1990–2012, no clear temporal trend in cyclone frequencies is found in the oceans around either Australia or China (Fig. 1). Nevertheless, during the very strong El Niño events in 1997–1998, cyclones were frequent in the oceans around Australia. However, no similar temporal trend is found in the ocean around China. Similarly, no clear temporal trend could be observed for cyclone intensities (i.e. minimum central pressure) in the oceans around either Australia or China during 1990–2012, even combined with El Niño intensities (Fig. 1).

We conducted regression analysis to examine the relationship between TD/GDP and wetland area, wind speed, the duration and minimum central pressure of cyclones for both countries, but non-significant
Figure 2. The relationship between TD/GDP and wetland area in cyclone swaths and/or the minimum central pressure of cyclones in Australia (a) and China (b). The regression equation in (a) $ln (\text{TD/GDP}) = -1.044 - 3.812 \times 10^{-3} \times \text{wetland area in cyclone swath}$. The shaded area is 95% confidence intervals. The regression equation in (b) $ln (\text{TD/GDP}) = 123.7 - 1.475 \times 10^{-3} \times \text{wetland area in cyclone swath} - 18.52 \times ln (\text{minimum pressure})$. 

Discussion

The variation of cyclone frequencies and minimum central pressure reflects the patchiness in cyclone occurrence and intensities. The frequent occurrence of cyclones in the oceans around Australia in 1997–1998 reflects the accumulated energy in the Southern West Pacific Ocean during the very strong El Niño events. The lack of similar trends for minimum central pressure in the oceans around Australia during 1990–2012 is probably related to the interannual variability of sea surface temperatures (SST).

Our study sheds light on the effect of wetlands in reducing the economic damage by cyclones for both countries. The regression models show that TD/GDP decreases with increasing wetland area and minimum central pressure in China. This negative effect of wetland area on TD/GDP is in accordance with that of Costanza, et al., although the regression models derived in the two studies are different. The study by Costanza, et al. assessed the annual economic value of coastal wetlands in shielding hurricanes, taking into account maximum sustainable wind speed, while the current study uses the minimum pressure as an estimate of cyclone intensity. This study further demonstrates that wetlands, especially coastal wetlands, are crucial in attenuating the impact of cyclones. The $R^2$ values of the results for both countries (Australia: 0.59; China: 0.09) are lower than that of the analysis on USA (0.6). This suggests that (1) there may be other variables explaining the variation of TD/GDP in Australia and China, in comparison with USA; or (2) the uncertainty in...
cyclone damage measurement may explain part of the difference, in particular for China, which has limited avenues to precisely quantify cyclone damage in the 1990s.

As shown in Ma, et al.,15 seawalls have been established extensively along the coastline of China since 2010, while wetlands reclaimed for seawall construction had the fastest areal increases during 1991–2010. This means that seawalls have replaced natural wetlands and have been the main mitigation measure for cyclones after 2010. Since seawalls were being continuously built and were not complete until 2010, the estimated value of TD/GDP is calculated without the effect of seawalls that could not be included as an independent variable in the regression model, while the actual value is based on relatively complete seawalls. Nevertheless, no significant difference in TD/GDP is found between the estimated and actual values, suggesting that seawalls may not provide significant protection of the coastlines. In addition to indicating the insignificant effect of seawalls for cyclone mitigation, we have collected and analyzed benefit/cost ratio data to show other ecosystem services provided by ‘green infrastructure’ (i.e. wetlands) but not provided by ‘grey infrastructure’ (including seawalls). The benefit/cost ratios were calculated as the ratio of the benefits of ecosystem services and the costs arising from service provision, e.g. operation and management. Wetlands provide other important ecosystem services, including wastewater treatment, recreation, carbon sequestration, flood control and fisheries (Table 2). The cost-benefit analysis reflects that wetlands have a high benefit/cost ratio (>5) for wastewater treatment and moderate benefit/cost ratio (>1) for recreation and carbon sequestration. No cost-benefit relationships were found for flood control and fisheries but both have high economic values ($971–35,034 ha⁻¹ yr⁻¹ for flood control and $2,805–17,090 ha⁻¹ yr⁻¹ for fisheries).

This result of TD/GDP for the cyclones in Australia and China is in agreement with the present (1981–2000) and future (2081–2010) baseline estimate of TD from tropical cyclones that TD in East Asia was significantly higher than that in Oceania. In particular, TD in China was one order of magnitude higher than that in Oceania for the future baseline estimate24.

The result of wetland value for cyclone mitigation in this study is comparable to previous studies evaluating the protective role of coastal wetlands for hurricane mitigation. Costanza, et al.,9 estimated MVₚ between $ 256 ha⁻¹

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**Figure 3.** Comparisons of TD/GDP in relation to the cyclones during 1990–2012 in China and Australia (a), and actual and estimated TD/GDP in relation to cyclones after 2010 in China (b). Bars labelled with different letters are significantly different from each other (Mann-Whitney test). Data are mean ± standard error.
Table 1. Annual expected MV per average cyclone swath and total value of wetland protection in relation to cyclones. Note: WA denotes Western Australia, NT denotes Northern Territory and QLD denotes Queensland.

| Ecosystem services               | Indices | Indices value          | References |
|----------------------------------|---------|------------------------|------------|
| Wastewater treatment             | Benefit | 2.0–6.5 (dollar-based)| Ko, et al. |
|                                  |         | 4.8–22.1 (energy-based)|            |
| Recreation and sequestration     | Benefit | 1.01–1.03              | Ghermandi and Fichtman |
| Flood control                    | Benefit | $971–35034 ha$^{-1}$ yr$^{-1}$ | Woodward and Wu; Camacho-Valdez, et al. |
| Fisheries                        | Benefit | $2805–17090.1$ ha$^{-1}$ yr$^{-1}$ | Woodward and Wu; Mukherjee, et al. |

Table 2. The cost-benefit/benefit analysis of important ecosystem services provided by wetlands in addition to cyclone mitigation (values were sourced from the stated references).
The annual expected MV per average swath has both similarities and differences to previous studies on the protective value of coastal wetlands. The differences may be due to the difference in value in various wetland types in cyclone mitigation. Barbier, et al. \(^6\) reviewed the value of coral reefs, salt marshes and mangroves for cyclone protection, and found that the value ranged from $174 ha^{-1} \text{yr}^{-1}$ to $10,821 ha^{-1} \text{yr}^{-1}$. In addition, these marine habitats are effective in dealing with different disasters\(^18,20\). For example, mangroves are most effective in attenuating storm-induced waves less than 6 m in height\(^32,33\), and the occurrence of a belt of mangroves of 100 m can attenuate wave height remarkably (6 m in height)\(^32,33\), owing to their complex aerial root structure\(^34\). Further, there are other factors contributing to the variation of protective value with reference to coastal wetlands, such as distances from the coast. The protection value tends to decrease with increasing distance onto land.

There are uncertainties and limitations in our estimates. Firstly, globally, cyclone damage is not well measured in all countries\(^24\) and this also applies to the countries evaluated herein. There is no international consensus in terms of best practices for collecting these data, and thus there remains considerable variability in reporting them. In particular, there is a lack of reported cyclone damages that hit Australia, including (1) cyclones reported with unknown damages; (2) cyclones reported with minor damages and unable to be quantified; and (3) anonymous cyclones, damage data of which cannot be found. This results in the patchy data in the regression analysis and some influence on our valuation of wetlands for cyclone mitigation. In particular, the probabilities of cyclone occurrence at intermittent categories in some states are zero, and may generate uncertainties for the valuation of wetlands in Australia. From the perspective of natural disaster management, Australia needs to strengthen the estimation of cyclone damages to facilitate effective disaster compensation after cyclone attack. Secondly, the impact of climate and human forces on cyclone damage could not be accounted for in our regression model. Strong interannual variability in cyclone statistics and the possible influence of interannual variability associated with E\(\text{I} \text{N} \text{i} \text{o} \text{ } \text{N} \text{e} \text{r} \text{s} \text{a} \text{d} \text{e} \text{s} \) events make it difficult to discern any temporal trend relative to background SST increases with correlation accuracy\(^26\). Consequently, any variability of SSTs associated with E\(\text{I} \text{N} \text{i} \text{o} \text{ } \text{N} \text{e} \text{r} \text{s} \text{a} \text{d} \text{e} \text{s} \) events could not be explained in the model for Australia. On the other hand, precise geographic information about seaways during the study period (1990–2012) could have further improved the model for China, but these data were unavailable. Finally, the weak regression model accounts for some uncertainties in the comparison of actual versus estimated TD/GDP, and the valuation of wetlands in China. Coastal wetlands have clear zonal compositions in different geographic regions of China, i.e. both mangroves and saltmarshes occur in subtropical regions but only saltmarshes in temperate and cold regions. Mangroves are trees while saltmarshes consist of grasses, sedges, bushes or other herbs. The physio-morphological difference of vegetation in the two types of coastal wetlands may explain part of the differences in the damage resulting from cyclones, but this is outside the scope of this study. Future studies are expected to improve the precision of wetland valuation if the area of different coastal wetland types can be quantified for each year of the study period.

**Methods**

We followed the approach of Costanza, et al.\(^8\) who estimated the value of wetlands in mitigating hurricanes hitting the USA from the relationship between TD/GDP and influential variables.

**Data collection and collation.** Data on the cyclones were assembled from pertinent web sites. First, we collected available data on the main cyclones (maximum sustained wind speed >120 km h \(^{-1}\)) striking Australia (17 of the total of 231) and China (97 of the total of 300) since 1990. TD information for 17 cyclones in Australia and 97 cyclones in China that were regarded as ‘disasters’ was collected from [www.emdat.be](http://www.emdat.be), which provides an objective basis for vulnerability assessment and rational decision-making in disaster situations, including disaster-related economic damage estimates. TD was transferred to US dollars of the base year (2011) based on implicit price deflator from World Bank. Based on established classification schemes ([www.nhc.noaa.gov/climo](http://www.nhc.noaa.gov/climo)), cyclones are classified as tropical depression, tropical storm and hurricanes (including major hurricanes), depending on the maximum sustained wind speeds. Where reported damages were remarkably different from this website, the cyclones were removed from our analysis, such as cyclone Yasi in 2011. Second, tracks of all cyclones hitting Australia and China from 1990–2012 were collected from [www.bom.gov.au](http://www.bom.gov.au) and relevant...
with the result of Costanza, from cyclones of a given GDP as well as the wetland area in swath. The total expected MV were calculated by cyclone swaths. The total expected MV and the unit average annual MV were derived from the regression analysis cyclone swaths were produced with the buffer function in geographic information systems, i.e. the average cyclones in specific categories (namely 1 to 5). Then a buffer zone of 100 km inland was created and 1 tracks were collected. They are critical to deriving a surrogate for the annual probability of wetlands being hit by cyclones. If seawalls can mitigate cyclones, the actual damage can reflect the effect, while the estimated damage definitely excludes the seawall influence which is not an explanatory variable in the regression model. By comparing the actual damage and estimated damage (from the regression analyses between TD/GDP and independent variables) after 2010, the effect of seawalls can be accounted for coarsely.

The effect of seawalls on the TD from cyclones in China cannot be directly evaluated since the geographic information of seawalls constructed in each year during 1990–2012 is not accessible. Therefore, the effect of seawalls accounts for part, if not all, of the variability in the relationship between TD/GDP and the independent variables but could not be accounted for due to the lack of geographic information. The seawalls were continuously being built during the study period but were incomplete before 2010. From 2010, seawalls were established along the China's coastline, and presumably effective, could account for the full strength of damage attenuation in cyclones. If seawalls can mitigate cyclones, the actual damage can reflect the effect, while the estimated damage definitely excludes the seawall influence which is not an explanatory variable in the regression model. By comparing the actual damage and estimated damage (from the regression analyses between TD/GDP and independent variables) after 2010, the effect of seawalls can be accounted for coarsely.

The collated data were combined to generate the values that are useful in our analysis. First of all, 100 × 100 km cyclone swaths were produced with the buffer function in geographic information systems, i.e. the average cyclone swath, and then overlaid on the wetland cover and GDP to produce GDP and wetland area in each cyclone swath. The total expected MV and the unit average annual MV were derived from the regression analysis from cyclones of a given GDP as well as the wetland area in swath. The total expected MV were calculated by multiplying the unit average annual MV and wetland area within 100 km of the coast. Subsequently, to compare with the result of Costanza, et al.8, data on storm frequency by state/territory and by pixel from historical cyclone tracks were collected. They are critical to deriving a surrogate for the annual probability of wetlands being hit by cyclones in specific categories (namely 1 to 5). Then a buffer zone of 100 km inland was created and 1 × 1 km pix- els were generated1. Data on the frequencies of cyclones hitting each administrative region (state/territory/prov- ince/municipality) were collected and used to estimate the annual probability of the pixel being hit by cyclones of specific categories. These probabilities were used to estimate damage by pixel, along with GDP and area of wetlands in the pixel.

Statistical and spatial analysis. Multiple regression analyses were conducted to evaluate the relationship between TD/GDP, wetland area, wind speed, and the duration and minimum central pressure of cyclones. Before regression analyses, data were checked for normality by the Shapiro-Wilk test ($\alpha = 0.05$). Significant outliers were removed via Cook’s distance. Variables were log-transformed where data violate the normality assumption. Data were also checked for possible interactions among independent variables by the variance inflation factor36. Step-wise regression analyses were performed to select the best model for predicting TD/GDP using the independent variables. We used the AIC to test alternative models, and the model with the lowest AIC from an array of models was selected. Paired-sample t-test was performed to compare the actual and estimated TD/GDP in China after 2010. Mann-Whitney test was used to compare TD/GDP between Australia and China, due to the unbalanced numbers of samples. Functions of spatial analyses, such as buffer, aggregate/disaggregate, rasterize, extract and mask, were used to conduct spatial analysis.

All the analyses were conducted by R programming language37. The R packages ‘raster’38, ‘rasterVis’39, ‘rgdal’40, ‘sp’41, ‘rgdal’40, ‘rgdal’40, ‘rasterVis’39, ‘geos’42, ‘colorspace’43, ‘maps’44 and ‘maptools’45 were used in the spatial analysis. The R package ‘rock- chalk’46 was used to produce the 3D plot of multiple regression analyses. The R package ‘relaimpo’47 was used to calculate the variance explained by minimum central pressure and wetland area in the regression model of China. Part of the computing task was assisted by the third-party application of R in the 79-core high performance computer cluster Gowonda (Intel Xeon CPU X5650 processor@2.67 GHz, QDR 4 × InfiniBand Interconnect) under Linux environment.

Data availability. Data are available upon request to oyxiaoguang@gmail.com.
References

1. Mitsch, W. J. & Gosselink, J. G. (2007). *Wetlands* (4th Edition). (John Wiley & Sons, Inc., 2007).
2. Ouyang, X. & Guo, F. (2008). Paradigms of mangroves in treatment of anthropogenic wastewater pollution. *Sci. Total Environ.* 344, 971–979. https://doi.org/10.1016/j.scitotenv.2006.01.066 (2008).
3. Ouyang, X. & Lee, S. Y. (2017). Updated estimates of carbon accumulation rates in coastal marsh sediments. *Biogeosciences* 11, 5057–5071. https://doi.org/10.5194/bg-11-5057-2014 (2014).

4. Wagelscheren, L., Sheaves, M., Baker, R. & Connolly, R. M. (2017). The seascape nursery: a novel spatial approach to identify and manage nurseries for coastal marine fauna. *Fish and Fisheries* 16, 362–371 (2015).
5. Nemerov, N. L. (2007). *Industrial Waste Treatment: Principles and Practice*. Butterworth-Heinemann, Oxford.

6. UNEP. (2006). Marine and coastal ecosystems and human well-being: a synthesis report based on the findings of the Millennium Ecosystem Assessment. UNEP (2006).

7. McInerney, G., Bald, D. & Anderson, B. (2016). The rising tide: assessing the risks of climate change and human settlements in low elevation coastal zones. *Environ. Urban.* 19, 17–37 (2007).

8. Costanza, R. et al. (2007). The value of coastal wetlands for hurricane protection. *AMBIO: A Journal of the Human Environment* 37, 241–248 (2008).

9. Spalding, M. D. et al. (2014). The role of ecosystems in coastal protection: Adapting to climate change and coastal hazards. *Ocean Coast. Manag.* 90, 50–57 (2014).

10. Lee, S. Y. et al. (2014). Ecological role and services of tropical mangrove ecosystems: a reassessment. *Glob. Ecol. Biogeogr.* 23, 726–743. https://doi.org/10.1111/geb.12152 (2014).

11. Titus, J. G. et al. (2012). State and local governments plan for development of most land vulnerable to rising sea level along the US Atlantic coast. *Environ. Res. Lett.* 7, 044008 (2012).

12. Rosenzweig, C. et al. (2012). Developing coastal adaptation to climate change in the New York City infrastructure-shed: process, approach, tools, and strategies. *Clim. Chang.* 106, 93–127 (2011).

13. Brown, S., Barton, M. & Nicholls, R. (2012). Coastal retreat and/or advance adjacent to defences in England and Wales. *J. Coast. Conserv.* 15, 659–670 (2011).

14. Stancheva, M. et al. (2012). Expanding level of coastal armouring: case studies from different countries. *J. Coast. Res.* 1815 (2011).

15. Ma, Z. et al. (2012). Rethinking China's new great wall. *Science* 346, 912–914 (2014).

16. Venkatachalam, A. J., Price, A., Chandrasekaran, S. & Senaratna Sellamuttu, S. (2009). Risk factors in relation to human deaths and other tsunami (2004) impacts in Sri Lanka: the fishers’-eye view. *Aquatic Conservation: Marine and Freshwater Ecosystems* 19, 57–66 (2009).

17. Chang, S. E. et al. (2004). Coastal ecosystems and tsunami protection after the December 2004 Indian Ocean tsunami. *Earthquake Spectra* 22, 863–887 (2006).

18. Das, S. & Vincent, J. R. (2009). Mangroves protected villages and reduced death toll during Indian super cyclone. *R. Natl. Acad. Sci. USA* 106, 7357–7360 (2009).

19. Pinsky, M. L., Guannel, G. & Arkema, K. K. (2013). Quantifying wave attenuation to inform coastal habitat conservation. *Ecosphere* 4, 1–16 (2013).

20. Wolanski, E. (2013). Coastal protection in the aftermath of the Indian ocean tsunami: what role for forests and trees? In: Braatza, S., Fortuna, S., Broadhead, J., Leslie, R. (Eds), Proceedings of an FAO Regional Technical Workshop, Khao Lak, Thailand, 28–31 August 2006. FAO, Thailand, pp. 157–179 (2007).

21. Nicholls, N., Lavery, B., Frederiksen, C., Drosdowsky, W. & Torok, S. (2013). Recent apparent changes in relationships between the El Niño-Southern Oscillation and Australian rainfall and temperature. *Geophysical Research Letters* 33, 3357–3360 (2012).

22. Hamilton, S. E. & Casey, D. (2013). Creation of a high spatio-temporal resolution global database of continuous mangrove forest cover for the 21st century (CGMFC-21). *Glob. Ecol. Biogeogr.* 25, 729–738 (2013).

23. Mcown, C. J. et al. (2013). A global map of saltmarshes. *Biodiversity data journal* e11764. https://doi.org/10.5194/bdij-e11764 (2017).

24. Mendelsohn, R., Emanuel, K., Chonabayashi, S. & Bakkensen, L. (2013). The impact of climate change on global tropical cyclone damage. *Nat. Clim. Chang.* 2, 205–209 (2012).

25. Barbier, E. B. (2013). The protective service of mangrove ecosystems: A review of valuation methods. *Mar. Pollut. Bull.* 69, 678–692 (2013).

26. Webster, P. H., Holland, G. J., Curry, J. A. & Chang, H.-R. (2013). Changes in tropical cyclone number, duration, and intensity in a warming environment. *Science* 309, 1844–1846 (2005).

27. Gedan, K. B., Kirwan, M. L., Wolanski, E., Barbier, E. B. & Silliman, B. R. (2013). The present and future role of coastal wetland vegetation in protecting shorelines: answering recent challenges to the paradigm. *Clim. Chang.* 106, 7–29 (2011).

28. Barbier, E. B. (2011). A spatial model of coastal ecosystem services. *Ecol Econ* 78, 70–79. https://doi.org/10.1016/j.ecolecon.2012.03.015 (2012).

29. Tam, W. et al. (2011). A historical typhoon database for the southern and eastern Chinese coastal regions, 1951 to 2012. *Ocean Coast. Manag.* 108, 109–115 (2015).

30. Zuur, A. F. et al. (2007). *Statistics for biology and health* (Springer, 2007).

31. R. (2014). *A language and environment for statistical computing*. Vienna, Austria (2014).

32. Hijmans, R. J. (2016). *Geographic Data Analysis and Modeling*. R package version 2.5–8. https://CRAN.R-project.org/package=raster (2016).

33. Lamnguero, O. P. & Hijmans, R. (2016). *meteo Forecast*. R package version 0.40 (2016).

34. Bivand, R., Keitt, T. & Rowlingson, B. (2016). *rgdal: Bindings for the Geospatial Data Abstraction Library*. R package version 1.1–10. https://CRAN.R-project.org/package=rgdal (2016).

35. Peiasma, E. J. & Bivand, R. S. (2016). Classes and methods for spatial data in R. *R News* 15, 67–70 (2016).

36. Bivand, R. & Rundel, C. (2016). *rgos: Interface to Geometry Engine - Open Source (GEOS)*. R package version 0.3–19. https://CRAN.R-project.org/package=rgos (2016).

37. Ihaka, R., Murrell, P., Hornik, K., Fisher, J. C. & Zeileis, A. (2016). *colorspace: Color Space Manipulation*. R package version 1.2–6. URL http://CRAN.R-project.org/package=colors (2016).

38. Minka, T. P. & Deckmyn, A. (2016). *A maps: Draw Geographic Maps*. R package version 3.1.0. https://CRAN.R-project.org/package=maps (2016).

39. Bivand, R. & Lewin-Koh, N. (2016). *maptools: Tools for Reading and Handling Spatial Objects*. R package version 0.8-39. https://CRAN.R-project.org/package=maptools (2016).
46. Johnson, P. E. rockchalk: Regression Estimation and Presentation. R package version 1.8.101. https://CRAN.R-project.org/package=rockchalk (2016).
47. Grömping, U. Relative importance for linear regression in R: the package relaimpo. J. Stat. Softw. 17, 1–27 (2006).
48. Ko, J.-Y., Day, J. W., Lane, R. R. & Day, J. N. A comparative evaluation of money-based and energy-based cost–benefit analyses of tertiary municipal wastewater treatment using forested wetlands vs. sand filtration in Louisiana. Ecol. Econ 49, 331–347 (2004).
49. Ghermandi, A. & Fichtman, E. Cultural ecosystem services of multifunctional constructed treatment wetlands and waste stabilization ponds: Time to enter the mainstream? Ecol. Eng. 84, 615–623, https://doi.org/10.1016/j.ecoleng.2015.09.067 (2015).
50. Woodward, R. T. & Wui, Y.-S. The economic value of wetland services: a meta-analysis. Ecol. Econ 37, 257–270 (2001).
51. Camacho-Valdez, V., Ruiz-Luna, A., Ghermandi, A. & Nunes, P. A. L. D. Valuation of ecosystem services provided by coastal wetlands in northwest Mexico. Ocean Coast. Manag. 78, 1–11, https://doi.org/10.1016/j.ocecoaman.2013.02.017 (2013).
52. Mukherjee, N. et al. Ecosystem service valuations of mangrove ecosystems to inform decision making and future valuation exercises. PloS one 9, e107706 (2014).

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Author Contributions
X.O. designed the study. S.Y.L. provided feedback on the framework of the study. X.O. collected and analysed the data. X.O., M.J.K., S.Y.L. and R.C. contributed to write the manuscript.

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