Hydrogeochemical characterization of groundwater from semiarid region of western India for drinking and agricultural purposes with special reference to water quality index and potential health risks assessment

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Received: 15 April 2020 / Accepted: 24 July 2020 / Published online: 7 August 2020
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Abstract
Increase in demand for water for domestic, agricultural and industrial needs has led to overexploitation as well as contamination; hence, it is unremittingly important to monitor and maintain the health of waters. In view of this, a study was carried out wherein a total of 57 groundwater samples from Upper Bhima Basin of western India were collected and analysed for understanding the drinking and agricultural suitability with special reference to water quality index and potential health risks assessment. Analysed physicochemical parameters indicated that the groundwater quality is not suitable in most of the villages for drinking and agricultural purposes. However, pH, TDS, HCO3−, SO42−, Cl− and F− parameters were below the permissible limit; conversely, Ca2+, Na+, Mg2+ and NO3− contents have been found more than the Indian standards (BIS). The dissemination of ions in the qualitative analysis revealed the dominance order as Ca2+ > Na+ > Mg2+ > K+ and HCO3− > SO42− > Cl− > NO3−. Using Piper’s diagram, the groundwater types were found to be majorly Ca2+–HCO3−, Ca2+–Na+–HCO3−, Ca2+–Mg2+–Cl− and Ca2+–Cl− groups. From the Gibbs plot, it is observed that the samples fall between rock dominance and evaporation dominance zones. Based on irrigation indices such as permeability index, residual sodium carbonate, the groundwater is unsuitable for irrigation. In addition to that, water quality index values revealed that 49% of water samples are found unfit for drinking. Further, health risk assessment study carried out using United States Environmental Protection Agency (US EPA) guidelines showed that children are at higher health risk than men and women in the study region.

Keywords Upper Bhima Basin · Water quality index · Health risk assessment · Semiarid · Western India

Introduction
Groundwater is the main source that caters the need of inhabitants from rural part for daily activities, especially in a developing country like India (Gaikwad et al. 2018; Kumar et al. 2020; Kate et al. 2020). All the agricultural practices carried out in the country are majorly dependent on this hidden natural resource (Srivastava and Parimal 2020; Marghade 2020; Verma et al. 2020). Specifically, in arid and semiarid provinces, subsurface water is mostly limited due to the scanty precipitation, high evaporation and surface runoff (Camacho Suarez et al. 2015; Kadam et al. 2020). While meeting the demands from various sectors, this precious resource is getting scarce, overexploited and contaminated in many parts of the country (Avannavar and Shrihari 2008; Vasanthavigar et al. 2010; Sharma...
and Kansal 2011). Moreover, the ingestion of such polluted groundwater could upsurge the hazards of aquatic infection epidemic (Wu and Sun 2016; Li et al. 2019). The degrading groundwater quality has become a serious universal issue for sustainable development (Li et al. 2019; Adimalla and Wu 2019). Hence, it is extremely important to understand the hydrogeochemical characteristics of groundwater for sustainable resource development and governance.

Various studies have been reported regarding the geochemistry of waters from semiarid region including the drinking and irrigation water suitability (Avvannavar and Shrihari 2008; Karyab et al. 2019; Mukate et al. 2019; Adimalla et al. 2020; Verma et al. 2020; Wagh et al. 2020), water quality index and health risk assessment (Li et al. 2016; Kavo and Karuppnan 2018; Mgbenu and Egbueri 2019; Adimalla et al. 2019; Zhang et al. 2020). It has been observed that in most of the studies, groundwater quality is deteriorated because of the upsurge in water demand due to fast-growing inhabitants as well as industrial development. Adimalla et al. (2018) evaluated the groundwater quality in Peddavagu area from Central Telangana (PCT), South India, for fluoride enrichment by rock–water interaction. Quijano et al. (2017) assessed the risk of nitrates in the Region of Valencia (Spain) and found that the exposures to nitrate from vegetables are improbable. Mgbenu and Egbueri (2019) examined the hydrogeochemical signatures, computed water quality indices and assessed health risk in the region of southeast Nigeria. The assessment results showed that the variations in hydrogeochemical signatures and quality of waters are attributed to the collective effects of man-made actions and lithogenic processes such as silicate weathering and ionic exchanges. Bamousa and El Maghraby (2016) have carried out the groundwater characterization and quality assessment to identify the sources of pollution in Madinah, Saudi Arabia. Deshmukh (2013) studied the impact of man-made activities on the quality of groundwater from Sangamner Area, Ahmednagar District, Maharashtra, India. Chidambaram et al. (2011) attempted the hydrochemical assessment of coastal groundwater in Southern India for major ions with the emphasis on fitness of groundwater for drinking and agricultural purposes. However, Avvannavar and Shrihari (2008) evaluated water quality index for domestic purposes in Nethravathi River basin from South India.

Subsequently, hydrogeochemical characterization of groundwater for domestic and irrigation purposes using water quality index and potential health risks assessment have been carried out by many researchers (Narsimha 2018; Zhang et al. 2018; Li et al. 2019; Adimalla 2020), as these are the most prevalent approaches to identify the groundwater suitability for different purposes. The studies carried out by various researchers (Panaskar et al. 2017; Mukate et al. 2018; Varade et al. 2018; Kadam et al. 2019a, b; Wagh et al. 2020) in Deccan Volcanic Province have shown that the natural factors such as geology, hydrodynamics and hydrogeomorphology along with man-made activities have brought major changes in chemical composition of groundwater. As the study region incorporates the rural areas, agriculture is the dominant source of income for the villagers. The source of groundwater for agriculture in this region is large-diameter dug wells, cost-efficient deep bore wells and in few places dug-cum-bore wells. In order to intensify the crop production, most common practices such as application of high amount of chemical fertilizers and pesticides are in vogue that leach downward and lead to contamination of water.

GIS is one of the most efficient techniques for evaluating the hydrogeochemical characteristics and the quality of groundwater, which has been implemented by many workers (Umrikar and Devi 2011; Wagh et al. 2019; Kadam et al. 2019a, b; Adimalla and Taloor 2020). Groundwater quality zonation map is helpful in identifying the contaminated locations of groundwater and also in indicating the potential hazardous zones that can assist various officials working in that region. In the present study, hydrogeochemical characterization of groundwater from semiarid region of western India is performed which assesses the suitability for domestic and agriculture purposes. Additionally, the groundwater suitability zones for domestic and agriculture purposes have been created using GIS technique. Finally, the health risk is assessed for children and adults (men and women) due to ingestion of water contaminated with high nitrate concentration. With the help of this study, improved observing plan of groundwater quality and preventive measures for susceptible zones would be implemented by administrative planners and decision-makers.

**Study area**

Upper Bhima Basin (Survey of India Toposheet No 47 J/5, 47 J/6, 47 J/9 and 47 J/10) area lies between 18° 30′ 00″ and 19° 00′ N latitudes and between 74° 10′ 00″ and 74° 30′ 00″ E longitudes (Fig. 1). The study area is comprised of four subwatersheds (BM-16, BM-27, BM-28 and BM-34) and is 810 km² in extent.

The highest elevation in the area is 791 m, and the lowest elevation is 425 m above the mean sea level. The area is about 70 km from Pune and is well accessible by road (Pune–Ahmednagar highway). The study area receives rainfall during early parts of June right up to the end of September that is very uncertain and erratic with an average annual of 497 mm. The average maximum temperature is about 32 °C and minimum 21 °C. The highest temperature in the summer months of April and May goes up to 40 °C representing semiarid climate of Western India; hence, evaporation is the dominant process.
The Upper Bhima Basin area is covered by basalt flows of Cretaceous–Paleocene age with Quaternary alluvial and colluvio-alluvial deposits occurring in the valleys. The lava pile exposed in the area is made up of both compound pahoehoe flows and a`a flows and has been grouped into Ratangarh and Indrayani Formations of the Sahyadri Group of Deccan Trap Super Group. These formations are separated by marker horizons Megacryst M3 and Giravalli GPB (Godbole et al. 1996; Hooper et al. 1988). Megascopically, the flows are light to dark grey, fine to coarse grained, nonporphyritic.
to porphyritic. The porphyritic ones are mostly plagioclase phyric, with fine- to coarse-size phenocrysts exhibiting sparsely to moderately porphyritic texture. Jointing is common in the dense part of the flows. The pahoehoe flows exhibit fewer regular joints as compared to a’a flows. They exhibit sheet joints towards the top which are clearly seen in the well sections cutting the pahoehoe flows. The a’a flows are highly jointed, forming colonnade and entablature zones and more than one tiers. Fractures in the area generally trend NNW–SSE and NE–SW.

Materials and methodology

Sampling and analytical procedures

The hydrogeochemical assessment of groundwater for domestic and agricultural purposes, computing water quality index and potential health risks assessment from semi-arid part of Western India were performed by collecting 57 groundwater samples from dug wells up to 18 m depth during post-monsoon (December 2018) season, while in pre-monsoon most of the dug wells get dry. The groundwater samples were collected in a bottle pre-washed with distilled water of 1 litre capacity and dried before fieldwork. The on-field measurements were done for pH, electrical conductivity (EC) and total dissolved solids (TDS) by using Hanna HI 9811–5 handheld meter. Further, the groundwater samples were analysed for various hydrochemical parameters such as calcium (Ca²⁺), magnesium (Mg²⁺) and total hardness (TH) by using APHA standard method, while sodium (Na⁺) and potassium (K⁺) were determined by flame photometer (Systronics Model 130) and major anions [chloride (Cl⁻), bicarbonate (HCO₃⁻), sulphate (SO₄²⁻), nitrate (NO₃⁻) and fluoride (F⁻)] were determined by high-pressure ion chromatography (HPIC).

The major ion concentrations in analysed samples were plotted on the Piper diagram using AquaChem software (AquaChem v4.0). The irrigation suitability indices such as sodium adsorption ratio (SAR), permeability index (PI), Kelly’s ratio (KR), magnesium hazard (MH), sodium percentage (Na%) and residual sodium carbonate (RSC) were calculated using the following formula (Table 1).

Water quality index (WQI)

Keeping in view the complaints of local people regarding dug well waters during the fieldwork, a water quality index (WQI) was determined to assess the groundwater suitability for drinking and domestic purposes (Horton 1965; Ramakrishnalah et al. 2009). Detailed facts and procedure about the WQI have been given by Asadollahfardi (2015). Eleven parameters (TDS, pH, TH, HCO₃⁻, Na⁺, K⁺, Ca²⁺, Mg²⁺, Cl⁻, SO₄²⁻ and NO₃⁻) were considered for the calculation of WQI values using Bureau of Indian Standards (BIS). Further, the suitable weights were allocated to each groundwater parameter. The parameters such as TDS, EC, Cl, SO₄ and NO₃⁻ were assigned higher weights of 5, whereas HCO₃ was assigned lowest weights 1 depending on their importance in determination of water quality (Wagh et al. 2017; Adimalla et al. 2018; Kadam et al. 2019a, b). The relative weights (W3) (Table 2) were calculated for each parameter using Eq. (1). WQI values were calculated using Eqs. (2), (3) and (4)

\[ RW_i = \frac{AW_i}{\sum AW_i} \]  

where \( RW_i \) is the relative weight, \( AW_i \) is the assigned weight and \( n \) is the number of parameters.

The quality rating scale for each parameter is expressed with the following equation:

| Sr. no. | Chemical parameters | Weight (\( w_i \)) | Relative weight |
|---------|---------------------|-------------------|----------------|
| 1       | TDS     | 5                 | 0.122          |
| 2       | pH      | 4                 | 0.098          |
| 3       | TH      | 4                 | 0.098          |
| 4       | HCO₃⁻   | 1                 | 0.024          |
| 5       | Cl⁻     | 5                 | 0.122          |
| 6       | SO₄²⁻   | 5                 | 0.122          |
| 7       | NO₃⁻    | 5                 | 0.122          |
| 8       | Ca²⁺    | 3                 | 0.073          |
| 9       | Mg²⁺    | 3                 | 0.073          |
| 10      | Na⁺     | 4                 | 0.098          |
| 11      | K⁺      | 2                 | 0.049          |

\[ \sum w_i = 41 \quad \sum W_i = 1.00 \]
\[ q_i = \left(c_i/s_i\right) \times 100 \]  

(2)

where \( q_i \) is quality rating for \( i \)th parameter, \( C_i \) is concentration of \( i \)th chemical parameter of water sample (mg/l), \( s_i \) is permissible standard for \( i \)th parameter set by BIS.

Subindex (SI) for each parameter is calculated by following equation:

\[ (SI_i = RW_i \times q_i) \]  

(3)

where \( SI_i \) is subindex of \( i \)th parameter, \( q_i \) is rating based on the concentration of \( i \)th parameter.

Finally, WQI is proposed using the following equation:

\[ WQI = \sum SI_i \]  

(4)

**Health risk assessment (HRA)**

To assess the potential human health risk due to ingesting of contaminated groundwater, the United States Environmental Protection Agency (US EPA 1989) developed a novel index that has been widely used (Narsimha and Rajitha 2018; Adimalla et al. 2019; Kadam et al. 2019a, b; Karande et al. 2020). There are many case studies reporting the problems of health risk due to consumption of contaminated water in India and all over the world (Karyab et al. 2019; Quijano et al. 2017; Li et al. 2016; Narsimha and Rajitha 2018; Adimalla et al. 2019; Karande et al. 2020). The health issues mainly arise from the contaminated air, water and soil surrounding the environment; hence, it is necessary to assess these resources and identify the sources of pollution. The present study area is mainly having the agricultural land use as major practices; therefore, elevated content of nitrate in dug well waters is a primary concern, where people are mostly reliant on the groundwater resources for domestic and agricultural purposes. Hence, the HRA was determined for NO\(_3^-\) by using the following equation:

\[ \text{Intake}_{\text{oral}} = \left(Cw \times IR \times EF \times ED\right)/(BW \times AT) \]  

(5)

where \( \text{Intake}_{\text{oral}} \) is the mean daily dose intake by ingestion of groundwater [mg/kg/day]; \( C_{w} \) is the content of a NO\(_3^-\) (mg/L) in the groundwater; \( IR \) shows the ingestion groundwater (L/day); \( IR = 1.5 \ L/\text{day for adults and 0.7 L/day children}; EF represents the exposure frequency (days/year); \( EF = 365 \text{ days/year}; ED \) indicates the duration of exposure (years; \( ED = 30 \text{ years for adults and 12 years for children, obtained from the US 1989}.\) \( BW \) is the average body weight in kg (BW = 65 kg and 55 kg for males and females, correspondingly, and 18.5 kg for children (ICMR 2009); and \( AT \) denotes the average time of exposure (days; \( AT = 10,950 \text{ days for adults and 4380 days for children).} \)

Hazard quotient is calculated for the nitrate using Eq. (6)

\[ HQ_{\text{oral}} = \frac{\text{Intake}_{\text{oral}}}{Rfd_{\text{oral}}} \]  

(6)

where \( HQ \) is the hazard quotient and \( Rfd \) is the average reference dosage, respectively. For NO\(_3^-\) reference dose will be 1.6 mg/kg/day according to United States Environmental Protection Agency (US EPA 2012). If the HQ value is greater than 1 and it is considered noncarcinogenic risk, while the safe limit for \( HQ_{\text{oral}} \) is equal to or less than 1.

**GIS analysis**

The results obtained from analysis were used to prepare spatial variation maps. These maps were prepared using inverse distance-weighted interpolation tool in ArcGIS Software 10.4 (Adimalla et al. 2018). The results were assessed according to the drinking water quality standards suggested by Bureau of Indian Standards (BIS 2012). The base maps were prepared by the heads-up digitization using survey of India Toposheet, and district resources map was used to prepare the geology of area. The Garmin GPS points were imported in GIS environment to plot the dug well location map. The spatial variation map of parameters such as WQI, SAR, PI, KR, MH, Na% and RSC was prepared. Spatial variation maps were prepared individually for male, female and children in accordance with HRA. Further, they were summed up using raster calculator tool in ArcGIS.

**Results and discussion**

**Hydrogeochemical characterization of groundwater**

In order to identify the point source of contamination in the groundwater, spatial distribution information is extremely useful. Most of the villages in India, especially those located in hilly areas, have groundwater as the only source for drinking and domestic requirements. Hence, the quality of groundwater is of utmost importance to secure the rural health. Thus, the suitability assessment of groundwater quality has become mandatory in the current scenario. The groundwater samples from the study area have been compared with BIS standards considering both permissible and desirable limits, and the data are presented in Table 3.

The pH of water is a function of host rock chemistry and its hydrological properties, as well as the rainfall, climate type and land use (Bartram and Ballance 1996). pH is vital parameter because it affects the biological and chemical properties of the groundwater. The BIS has prescribed a permissible range of 6.5–8.5, and it is observed that the pH values in the groundwater samples range from 6.8 to 7.9, i.e. well within the permissible limits. The EC ranges from 418 to 2987 μs/cm showcasing higher values due to the natural contents of salts as evapotranspiration surpasses...
precipitation in a semiarid condition. TDS count is the sum of all ions that are dissolved in an aqueous state. TDS values range from 264.67 to 1916.81 mg/L. The maximum desirable limit of BIS is 500 mg/L according to BIS samples found in the southern part which have TDS values higher than the permissible limit making it a hot spot. This area is in the proximity of dam having lower elevation experiencing intense agricultural practices, which are the major contributing factors to such high TDS content. The result of TDS shows that 81% of samples have higher TDS value than BIS desirable limit of 500 mg/L, whereas all samples are below the BIS permissible limit of 2000 mg/L (Table 3). Hardness in the groundwater is a measure of total sum of calcium and magnesium ion concentrations. The TH values range between 186 and 960 mg/L. About 61% of the samples have higher TH values than the desirable limit of 300 mg/L and 21% above the permissible limit of 600 mg/L.

The calcium ion concentration in the groundwater ranges from 15.31 to 214.02 mg/L. The average concentration of calcium ion is 90.18 mg/L; it is well above as compared with BIS standard desirable limit of 75 mg/L. The Ca$^{2+}$ ion is the dominant cation of the study area. It is observed that 58% of the samples are above the desirable limit and only 2% of the samples are exceeding the permissible limits (Table 3). Higher values of calcium ion are found in the area where groundwater has higher residence time, which in turn results in high degree of rock–water interaction and CaCO$_3$ is precipitated in water (Pawar et al. 2008). Magnesium and calcium have the similar behavioural characteristic in water but in general magnesium is different than that of calcium in their geochemical characteristics. Magnesium has smaller ions as compared to calcium (Hem 1985). Having the similar behavioural pattern, the concentration of magnesium is less due to less general abundance. Magnesium ion concentration in the Upper Bhima Basin ranges from 9.16 to 126.70 mg/L. The permissible limit for magnesium is 100 mg/L (BIS), and only 2% of samples are found to be above the permissible limit. The groundwater of the study area is normally fit for consumption purpose with respect to magnesium. Main source of Na in the groundwater is weathering of rocks; however, higher values show that sodium ion also arises out of anthropogenic activities (Subbarao et al. 1996). Na concentration ranges from 4.92 to 375.73 mg/L. The maximum desirable limit for sodium ion is 200 mg/L according to BIS standard showing 19% samples are above the desirable limits. Normally, potassium ion concentration in natural water is less than sodium ion having the concentration less than 10 mg/L in the groundwater. The value for potassium ion varies from 0.01 to 2.56 mg/L as a result of weathering of host rocks and anthropogenic activities.

The dissemination of major ions analysed reveals the dominance as HCO$_3^-$ > SO$_4^{2-}$ > Cl$^-$ > NO$_3^-$ hierarchical order in Upper Bhima Basin. Bicarbonate ion is one of the dominant anions and key source of alkalinity, and the subsidence from carbonate weathering processes and reaction of silicates with carbonic acid account for addition of HCO$_3^-$ to groundwater (Chidambaram et al. 2011). The desirable limit of bicarbonate ion in drinkable water is 300 mg/L (WHO 2011). Analysis of groundwater samples shows that bicarbonate concentrations range from 129 to 490 mg/L, and out of the total samples, 26% are above the desirable limit. High concentration of bicarbonate ion is found in most of the samples from the lower part of the study area showing decline in water table due to overexploitation of groundwater for

### Table 3: Descriptive statistics of physicochemical analysis of groundwater samples

| Parameter $^a$ | Min   | Max   | Average | Maximum desirable limit (MDL) | Maximum permissible limit (MPL) | % of sample above MDL | % of sample above MPL |
|----------------|-------|-------|---------|--------------------------------|---------------------------------|-----------------------|-----------------------|
| pH             | 6.80  | 7.90  | 7.33    | 6.5–8.5                        | –                               | 0                     | 0                     |
| EC             | 418.00| 2987.00| 1349.14 | –                               | –                               | –                     | –                     |
| TDS            | 264.67| 1916.81| 857.39  | 500                             | 2000                            | 81                    | 0                     |
| TH             | 186.00| 960.00| 486.62  | 300                             | 600                             | 61                    | 21                    |
| Ca$^{2+}$      | 15.31 | 214.02| 90.18   | 75                              | 200                             | 58                    | 2                     |
| Mg$^{2+}$      | 9.16  | 126.70| 58.34   | 30                              | 100                             | 84                    | 2                     |
| Na$^+$         | 4.92  | 375.73| 92.80   | 200                             | –                               | 19                    | –                     |
| K$^+$          | 0.01  | 2.56  | 0.57    | 12                              | –                               | 0                     | –                     |
| HCO$_3^-$      | 125.00| 490.00| 272.29  | 200                             | 600                             | 81                    | 0                     |
| SO$_4^{2-}$    | 1.38  | 369.12| 150.85  | 200                             | 400                             | 25                    | 0                     |
| Cl$^-$         | 35.50 | 367.40| 154.94  | 250                             | 1000                            | 21                    | 0                     |
| NO$_3^-$       | 0.33  | 108.23| 32.56   | 45                              | –                               | 26                    | –                     |
| F$^-$          | 0.02  | 0.91  | 0.55    | 1                               | 1.5                             | 0.00                  | 0.00                  |

$^a$All values in mg/L except pH and EC (μS/cm)
irrigation purposes and water–rock interaction. Chloride is one of the most abundant anions which are found in almost all-natural groundwater resources in varying proportions as it has the property of being highly stable (Sawyer et al. 1978; Stites and Kraft 2001). However, feldspathoids such as sodalite, Na₈ [Cl₂ (AlSiO₄)₆] and the phosphate minerals such as apatite (Hem 1985) are some of the lithogenic sources occurring in igneous rocks. But in the present case, it seems highly unlikely, as the concentrations are tremendous for claiming the geogenic source. The Cl content in the groundwater ranges from 35.50 to 367.40 mg/L with the mean of 154.94 (Table 3). The study area has 21% samples above BIS desirable limit for drinking water, and all samples were found to be beyond the permissible limits of BIS standards. At some places, Cl is high, because of anthropogenic activities, mainly sewage discharge from the populated area. The metallic sulphides are the major source of sulphate in the groundwater (Tiwari et al. 2013). In agricultural fields, around half to two-thirds of proportion of water is lost by evapotranspiration and the remaining amount infiltrates and joins the subsurface as a return flow. This infiltration of water into aquifers increases the salinity of groundwater (Vetrimurugan and Elango 2015). The significant concentration of SO₄²⁻ ion has been found in the regions that are highly irrigated, also anthropogenic stresses acting in the region. The sulphate content in the groundwater ranges from 1.68 to 369.12 mg/L with the average value of 150.58 mg/L (Table 3). Sulphate ion concentration in the study area is enhanced by the agricultural practices as the chemical fertilizers are used for augmenting the nutrient contents of soil. In the study area, 25% of the samples have concentration more than 200 mg/L (BIS standard), which shows that the groundwater in the area is significantly affected by manmade actions, which might be the resultant of local sewages and composts. 

The emergence of nitrate ion in our subsurface natural resource completely relies upon the lightening experiences during pre-monsoon thunder showers, biological nitrogen fixation by microbes (e.g. Rhizobium bacteria) and nitrogenous fertilizers application during crop cultivation period after monsoon season. The water-soluble nitrate fertilizers (e.g. urea, ammonium sulphate) easily get accumulated into groundwater system, and excess of these result in serious nitrate pollution (Pawar and Shaikh 1995). The nitrate (NO₃⁻) ion is the highest responded type of nitrogen in the atmosphere and is moderately nontoxic to amphibians (Vetrimurugan and Elango 2015). The high content of nitrate along with phosphate results in eutrophication as the main environmental problem. The NO₃ in the study area ranges from 0.33 to 108.23 mg/L (Table 3). Actually, various studies have detected a high relationship among farming practices and nitrate content in subsurface and surface waters (Adimalla et al. 2018). The contamination of nitrate ions is not only found in India but also observed globally (Margarhade 2020; Li et al. 2019; Karyab et al. 2019; Quijano et al. 2017; Majumdar and Gupta 2000). Many environmental scientists/workers have classified nitrate sources into a nonpoint type such as chemical fertilizers and point sources such as septic tanks and sewage systems (Narsimha and Sudarshan 2013; Adimalla 2020; Rao 2002). About 26% of the samples are above the BIS desirable limit of 45 mg/L. Classification of human health risk (Table 4) shows that the nitrate concentration in 26% and 2% of groundwater samples have high and very high risks, respectively. High nitrate concentration, above 45 mg/L, can cause the methaemoglobinaemia development and gastric and/or oesophageal cancer to exposed inhabitants (Adimalla et al. 2019).

Fluoride is found at higher concentrations in the groundwater, especially in arid and semiarid areas in different geological environments in the world, where the countryside inhabitants grieve from fluorosis disease, since there are no sources of water for drinking. However, fluoride concentration ranges from 0.02 to 0.91 mg/L, with a mean value of 0.56 mg/L (Table 3), and is found below the desirable limit (BIS 2012). About 25.71% of groundwater sampling locations presented risk for dental fluorosis, and 18.10% and 4.76% of samples indicated possibility for dental to skeletal fluorosis in the study region (Table 6).

### Groundwater types

The groundwater types of the study area were determined based on its geochemo characteristics of major ions; Piper (1953) proposed an effective plot. This is helpful to know the total element character of waters in terms of cations and anions composition. Using Piper diagram (Fig. 2), groundwater of the study area is categorized into: (I) Ca²⁺–HCO₃⁻, (II) mixed Ca²⁺–Na⁺–HCO₃⁻, (III) mixed Ca²⁺–Mg²⁺–Cl⁻ and (IV) Ca²⁺–Cl⁻ types. From the Piper plot, it is evident that the groundwater is mainly constituted of calcium, magnesium and bicarbonates. Calcium and magnesium impart

### Table 4 Nitrate and fluoride concentrations and health risk assessment

| Elements health risk | Concentration (mg/L) | Number of samples | % of samples |
|----------------------|----------------------|-------------------|-------------|
| Nitrate              |                      |                   |             |
| No risk              | ≤45                  | 41                | 72          |
| High risk            | 46–100               | 15                | 26          |
| Very high risk       | >100                 | 1                 | 2           |
| Fluoride             |                      |                   |             |
| Dental caries        | ≤0.5                 | 29                | 51          |
| Required levels for human health | 0.6–1.5 | 28 | 49 |
| Dental fluorosis     | 1.6–2                | 0                 | 0           |



hardness to the rock and are found due to weathering processes which lead to dissolution of minerals from parent rock, i.e. basalts (Gaikwad et al. 2020). Bicarbonates, on the other hand, are dominant due to weathering of carbonates and dissolution of silicate minerals. Higher values are also indicative of the anthropogenic activities. According to Piper, there is not much dominance of sodium–chloride–sulphate in the groundwater samples.

**Gibbs diagram for geochemical evolution**

In geochemical evolution of groundwater, there are variegated proceedings involved. To understand the dominance of either precipitation, rock matter or evaporation process, Gibbs (1970) designed two semilog diagrams which are now known as the Gibbs diagrams, in which graphs of TDS versus $\frac{\text{Cl}^-}{(\text{Cl}^- + \text{HCO}_3^-)}$ and TDS versus $\frac{(\text{Na}^+ + \text{K}^+)/}{(\text{Na}^+ + \text{K}^+ + \text{Ca}^{2+})}$ for anions and cations are plotted, respectively (Fig. 3). From the plot, it is evident that the samples fall between rock dominance zone and evaporation dominance zone. As the general slope in the region is low, there is an increase in residence time of groundwater. This results in high rock–water interaction. As the groundwater is from semiarid climatic region, evaporation dominance is obvious.

Evidences of evaporation dominance have been observed in the well sections, where the walls were covered with calcrete formation. As a result of high residence time, rock–groundwater interaction is observed in these basaltic aquifers.

**Water quality index (WQI) for domestic suitability**

Quality of water is very significant to human because it has a direct link with human health and welfare. Based on the WQI values obtained, the categories are generated as very poor (200–300), poor (100–200), good (50–100) and excellent (≤ 50). Out of the total of 57 samples, 27 are found very poor, one poor, 25 are good and only four are of excellent quality. Hence, it is inferred that 49% of water samples are found unfit for drinking. The spatial variation map (Fig. 4) shows that very poor quality is in the lower reaches of water-shed, while rest of the area has poor to good water quality status.

**Irrigation water quality**

There should be a periodic check for quality of water that is used for crop growth. It helps to increase the crop yield and supports farming of superior crop varieties. With the
climate change, dependency on rainwater for farming is uncertain. Hence, the suitability assessment of groundwater for agriculture is important task from planning perspective. Essential parameters comprise numerous definite properties of water related to the quality and quantity of crops and conservation of soil efficiency. Therefore, the groundwater suitability indices were computed for irrigation. Suitability indices such as sodium adsorption ratio (SAR), permeability index (PI), Kelly’s ratio (KR), magnesium hazard (MH), sodium percentage (%Na) and residual sodium carbonate (RSC) are discussed below:

**Sodium adsorption ratio (SAR)**

The sodium hazard is calculated by the absolute and relative concentration of cations in the groundwater and is termed as SAR (Richards 1954). The SAR value is the sodium content that exists in soil after absorption.

Higher values of SAR lead to high sodium concentration resulting in reduction in porosity of soils. The SAR values for most of the samples are found to be excellent, and those for few samples are unsuitable. About 58% of samples were found to be excellent, 16% of samples were good, 5% of samples were doubtful and 21% of samples were unsuitable according to sodium hazard. Accordingly, it is inferred that 74% of samples had no salinity issue and 26% of samples had salinity; hence, some treatment is required before it can be used for agriculture purposes. But, if the SAR value surpasses 10, the groundwater can cause porousness problems in soil. The spatial variation in SAR values shows that northern part of the watershed covering Ralegan Siddhi village in Ahmednagar district, Shirur and Malthan major villages from Pune district has unsuitable to doubtful groundwater for irrigation (Table 5, Fig. 5). Majorly, the catchment part of dam is affected by high SAR values.
The suitability assessment can also be conducted by considering the permeability index method. The soil permeability is an important factor as it allows the infiltration of water in subsurface. The PI values are categorized as excellent, good and unsuitable based on the suitability of groundwater for the agricultural purposes. From the calculated PI values for all the water samples, it is observed that 11% fall under the class excellent, 42% class good and 47% class unsuitable. The spatial variation in PI in the groundwater of the area shows that Ralegan Siddhi village in Ahmednagar district reveals suitable water type (Table 5, Fig. 6). The unsuitable zones covering Chichani village are mainly present in lower reaches of watershed, mainly in command areas.

Kelly’s ratio (KR)

KR is also another important parameter that can be used assess the quality of groundwater used in irrigation by considering sodium content present in the groundwater. As compared with SAR, it is not much advanced method for the detection of sodium content. It has only two simple classes, namely suitable and unsuitable. KR values above 1 are called as unsuitable, while those less than 1 are considered as suitable for irrigation purposes. Greater values show high concentration of sodium present. Kelly index values for the groundwater from 57 locations depict that 68% of samples...
are suitable and 32% of samples are unsuitable for irrigation, suggesting their adversarial effects on agricultural activities. The spatial variation in KR in the groundwater of the area shows that Ralegan Siddhi village in Ahmednagar district and Shirur and Malthan major villages from Pune district have unsuitable water type for irrigation (Table 5, Fig. 7). According to the KR method, the groundwater from the upstream part of the study area is unsuitable for irrigation.

**Magnesium hazard (MH)**

Magnesium and calcium concentrations are in equilibrium in the groundwater and play a vital role in defining the quality type of groundwater. Elevated magnesium content has more adverse effects as compared to calcium on crop yield by converting the soil alkaline. Hence, it is extremely important to detect the amount of magnesium which is done with the help of magnesium hazard (MH) (Paliwal 1972). Based on this, the values obtained by calculation were grouped into two classes, suitable and unsuitable. Values less than 50 were considered as suitable, and values exceeding 50 were considered unsuitable. MH values for 57 samples depict 81% of samples are suitable and only 19% of samples are unsuitable for irrigation. The central part of the study area mainly shows unsuitable water type (Fig. 8). Higher occurrence of magnesium ion in agricultural waters decreases the soil fertility which changes it to alkaline and finally reduces crop yield.

**Sodium percentage (Na%)**

Sodium hazard in irrigation is calculated with the help of sodium percentage, thus being a very useful parameter. Sodium content has extremely adverse effects on soil. Na is the major content in soil and water because it has intense influence on soil penetrability and its structure (Rao 2002). Generally, Wilcox (1955) and Richards (1954) methods are used to understand the Na%. From the values calculated, 31 samples were excellent, five good, seven in permissible limits and 14 samples doubtful (Figs. 9, 10).

In Wilcox diagram (1955), the values of EC are plotted against Na%, and from this plot, it was inferred that the groundwater samples are observed in all the classes. According to the diagram, about 50% of samples fall in excellent–good–permissible class (Fig. 11).

There were a few sample locations which are permissible/doubtful to doubtful and hence need proper treatment before use. The higher content of sodium ions in the groundwater is due to the long contact period with host rock resulting in dissolution of minerals and also the use of chemical fertilizers (Adimalla et al. 2018).

**Residual sodium carbonate (RSC)**

The hazardous effect of carbonates and bicarbonates was calculated with the help of residual sodium carbonate (RSC), detecting the attributes of groundwater for irrigation.
purposes. Particularly, the high amount of feeble acids such as carbonates and bicarbonates in the groundwater over the alkaline earths such as the total of calcium and magnesium alters the irrigating applicability of the groundwater. Eaton (1950a, b) concluded that increasing amount of RSC values in the groundwater increases SAR values, thus resulting in the soil becoming infertile due to such increasing concentration levels. Results show that only 5% of samples are suitable and 95% are unsuitable, which require some treatment before use. The constant use of groundwater having high RSC decreases the yield of crops (Ramesh and Elango 2012) (Fig. 12).

Human health risk characterization

Nitrate ion concentration in the groundwater used for drinking may pose an adverse health risk when the levels surpass the maximum permissible limit of 45 mg/L (BIS 2012). The noncarcinogenic risk due to nitrate ions in the groundwater is the cause of main worry for several nations (Quijano et al. 2017; Adimalla and Qian 2019a, b; Li et al. 2019; Karyab et al. 2019; Marghade 2020), particularly in India, where most of the inhabitants rely on groundwater for drinking and domestic purposes. Characteristically, the groundwater is mainly contaminated by nitrate ion and it poses a severe health risk as it breaks in the body into nitrite, which affects oxygen transmission by tie-up with haemoglobin and results in methaemoglobinaemia, which is particularly life threatening to infants and children (Majumdar and Gupta 2000; Fytianos and Christophoridis 2004; Gupta et al. 2011; Quijano et al. 2017; Karyab et al. 2019; Marghade 2020). Taking this fact into consideration, the US EPA has recommended a HRA method. Also, several scholars all over the world have used the US EPA method to evaluate the health risk for adults (men, and women) and children (Li et al. 2016; Quijano et al. 2017; Narsimha and Rajitha 2018; Adimalla et al. 2019; Karyab et al. 2019; Karande et al.)
Hence, in this study, human health risk of adults (men and women) and children due to elevated NO$_3^-$ concentration in the groundwater has been predicted. Results of percentage of groundwater samples representing health risk to adults and children are presented in Table 6. The health risk assessment map was prepared for male, female and children separately in accordance with HRA analysis of above one and below one HRA value and reclassified. Further, they were summed up using raster calculator tool in ArcGIS. The spatial variation map of cumulative health risk assessment shows that in the lower reach part of study area, all contaminates get concentrated having higher risk (Fig. 13).

Most of the samples are found above 1 indicating high risk in adults as well as children in the study region. According to the integrated index, 33% of children are prone to high health risks, while the male and female are 9% and 16% (Table 6). Hence, from the risk assessment, children $\geq$ female $\geq$ male is in hierarchy of susceptibility to risks in the given study region. In India, there are maximum children that are suffering from various diseases such as dental fluorosis due to the consumption of contaminated groundwater.

**Conclusions**

The agriculture is a dominant occupation in the study area, and groundwater demand in future will be very high. Along with the concern on the availability of this natural resource, it becomes extremely important to monitor and maintain its compositional value for safeguarding human’s health while maintaining the fertility of soil. This study is focussed on understanding the geochemistry of groundwater, its spatial variation and degree of pollution.

- From the analysis, it is observed that groundwater in this region is of neutral to alkaline type. Analysed physicochemical parameters indicate that few of the samples are within allowable limits of the BIS. Moreover, most of the parameters exceed the desirable limit at some sampled locations. Nitrate ion content exceeds the recommended limit of drinking in many groundwa-

| HI   | Health risk | No of samples | % of samples |
|------|-------------|---------------|--------------|
| Male | < 1         | No risk       | 52           | 91           |
|      | > 1         | High risk     | 5            | 9            |
| Female | < 1         | No risk       | 48           | 84           |
|      | > 1         | High risk     | 9            | 16           |
| Children | < 1         | No risk       | 38           | 67           |
|      | > 1         | High risk     | 19           | 33           |
ter samples. A high content of TDS was observed due to accumulation of salt and anthropogenic inputs.

- Qualitative analysis shows that major cations have dominance of $\text{Ca}^{2+} > \text{Na}^+ > \text{Mg}^{2+} > \text{K}^+$, and in major anions, it is $\text{HCO}_3^- > \text{SO}_4^{2-} > \text{Cl}^- > \text{NO}_3^-$ order. Alkaline earths ($\text{Ca}^2+$ and $\text{Mg}^{2+}$) exceed the alkalics ($\text{Na}^+$ and $\text{K}^+$), and weak acids ($\text{HCO}_3^-$ and $\text{CO}_3^{2-}$) exceed strong acids ($\text{Cl}^-$ and $\text{SO}_4^{2-}$).

- From Piper's diagram, the groundwater facies identified in the region are (I) $\text{Ca}^{2+}$–$\text{HCO}_3^-$, (II) mixed $\text{Ca}^{2+}$–$\text{Na}^+–\text{HCO}_3^-$, (III) mixed $\text{Ca}^{2+}$–$\text{Mg}^{2+}$–$\text{Cl}^-$ and (IV) $\text{Ca}^{2+}$–$\text{Cl}^-$, which represent host rock contribution and anthropogenic inputs. Gibbs plot shows that groundwater in this region is of rock dominance and evaporation dominance types. Calcium, magnesium, sodium and bicarbonates are found to be dominant in the groundwater samples. Calcium and magnesium impart hardness to the subsurface natural water resource that has geogenic source. Moreover, the sodium and bicarbonates are derived from anthropogenic activities.

- It was observed that in the study region, farmers apply low-grade fertilizers to enhance the crop productivity. Major cause of concern about these low-grade fertilizers is that they are rich in nitrate concentrations. The maximum permissible limit of nitrate in the groundwater is 45 mg/L, and about 28% are exceeding this limit. These samples are located in the region of intense agricultural practices.

- Based on water quality index, it was inferred that 49% of water samples are found unfit and 51% are suitable for drinking. The irrigation indices values of PI and RSC depict that groundwater is unsuitable for irrigation, while SAR, MH, KR and $\%\text{Na}$ show excellent and safe water quality.

- The study performed noncarcinogenic HRA considering the exposure of nitrate-contaminated drinking water in the rural region for men, women and children. The total noncarcinogenic risk ranged from 0.0048 to 1.56 for men, 0.0057–1.85 and 0.0079–2.56 for children and women, respectively. Thus, from the risk assessment, children $\geq$ female $\geq$ male is the order of decreasing susceptibility to health risks in the study region. The consumption of polluted groundwater is the main reason for the health risk, and thus the drinking water sources such as dug and bore well have to be monitored regularly to avoid the adverse affect on health. However, groundwater in some parts of the watershed is on the verge of getting polluted; hence, precautionary measures should be adopted in advance. In participatory groundwater movement, government officials, NGOs and farmers should come together and be educated to avoid any health complications arising due to water.

### Compliance with ethical standards

**Conflict of interest** The authors declare that they have no conflict of interest.

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