Plant cell (*Brassica napus*) response to europium(III) and uranium(VI) exposure

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Abstract

Experiments conducted over a period of 6 weeks using *Brassica napus* callus cells grown in vitro under Eu(III) or U(VI) stress showed that *B. napus* cells were able to bioassociate both potentially toxic metals (PTM), 628 nmol Eu/g fresh cells and 995 nmol U/g fresh cells. Most of the Eu(III) and U(VI) was found to be enriched in the cell wall fraction. Under high metal stress (200 μM), cells responded with reduced cell viability and growth. Subsequent speciation analyses using both metals as luminescence probes confirmed that *B. napus* callus cells provided multiple-binding environments for Eu(III) and U(VI). Moreover, two different inner-sphere Eu3+ species could be distinguished. For U(VI), a dominant binding by organic and/or inorganic phosphate groups of the plant biomass can be concluded.

Keywords Uranium · Europium · Plant cells · Luminescence spectroscopy · Viability

Introduction

The transfer of radionuclides such as actinides through the environment represents a critical safety concern for both nuclear waste repositories and former uranium mining and milling sites that must be made secure. Similarly, the potential harm associated with the growing use of lanthanides, for instance in high-tech products, has resulted in an elevated release of these elements into the environment, which may also represent severe health risk for humans. Due to the fact that lanthanides and actinides display many similarities based on their comparable ionic radii for elements of the same oxidation state and their analogous aqueous chemistry, lanthanides are considered to be suitable chemical analogs for actinides from a (bio-)chemical point of view. For instance, Eu(III) represents an analog for the trivalent actinides americium(III) and curium(III).

The accumulation of radionuclides and other potentially toxic metals (PTMs) into plants, and thus into the food chain, represents a potential pathway for human exposure. Plants need trace elements, e.g., copper and zinc, which represent important micronutrients for metabolic maintenance. However, in higher concentrations, all metal ions are toxic. Actinides and lanthanides, e.g., uranium and europium, are generally non-essential elements and are unlikely to have a special route for transport into plants; nonetheless, they can be taken up by plants and may interfere with normal metabolic processes. For instance, PTMs can replace essential metal ions from their binding sites in enzymes, damage sulphydryl-group-containing proteins, accelerate the formation of reactive oxygen species, and trigger antioxidant defense mechanisms in plants (e.g., Weiler and Nover 2008; Serre et al. 2019; Aranjuelo et al. 2014). To overcome this unwanted chain of events, plants synthesize protective metal binding metabolites, store metal chelates in vacuoles or secrete them into the rhizosphere (Weiler and Nover 2008), and deposit defense polymers such as callose or lignin (Serre et al. 2019).

The interaction of actinides and lanthanides with plants is often described in terms of transfer factors. In contrast, studies exploring the underlying mechanisms of these interactions at the cellular level, or those investigating toxic metal speciation at the molecular level, are less common. A recent study...
described the physiological and cellular responses of *Arabidopsis thaliana* roots to U stress (Serre et al. 2019); among several findings, the authors reported the deposition of the defense polymers callose and lignin in the roots due to uranium stress. Earlier studies of the interaction of uranium with plants revealed, for example, the importance of radionuclide speciation for the uptake and translocation of radionuclides in plants (e.g., Ebbs et al. 1998; Laurette et al. 2012a, 2012b), as well as the effects of uranium on phosphate homeostasis regulation (Misson et al. 2009; Berthet et al. 2018). In addition to the speciation effects on uranium uptake and the oxidative stress response (Saenen et al. 2013, 2015), the redox state of uranium and the influence of uranium on the intracellular glutathione pool of plants have also been investigated (Viehweger et al. 2011). The in situ speciation of uranium in plants (Günther et al. 2003) and their subcellular compartments (Geipel and Viehweger 2015) have been confirmed by spectroscopy. In a recent study, Sachs et al. (2017) combined isothermal microcalorimetry with spectroscopy and thermodynamic modeling to investigate the correlation between U(VI) toxicity in plant cells with oxidoreductase activity and U(VI) speciation. Earlier, Drake et al. (1997) used lanthanide ion probe spectroscopy in order to characterize the Eu$^{3+}$ binding sites on *Datura innoxia* cell wall fragments. Similarly, Eu$^{3+}$ uptake and partitioning on the common oat (*Avena sativa*) were investigated using time-resolved laser-induced fluorescence spectroscopy (TRLFS) and confocal microscopy profiling (Fellows et al. 2003). The authors confirmed the existence of Eu$^{3+}$ inner-sphere mononuclear complexes within the root. The impact of the europium speciation on its accumulation in *Brassica napus* and over-expressing *BnTR1* lines was studied by Zha et al. (2014).

The utilization of in vitro callus cell cultures represents an effective method for studying the physiological and biochemical response mechanisms to several stress factors at the cellular level (e.g., Huang et al. 2017a). Principally, callus cells are superior to the intact plant due to the simpler organization of their cells and tissues, thus augmenting the ability to more tightly control their growth conditions. Moreover, as discussed by Zagoskina et al. (2007), this approach also facilitates the ability to synthesize secondary metabolites that are characteristic of intact tissues. Callus cells have already been used to study the impact of PTMs on the growth of plant cell tissues. Maróti and Bognár (1989) investigated the growth inhibition of *Ruta graveolens* L. callus tissues in the presence of varying amounts of Cd, Cu, Hg, Ni, Pb, and Zn. Some years later, the effects of Cu on callus growth and the gene-expression of explants of *Nicotiana glauca* were reported by Taddei et al. (2007). The impact of Cu stress on the growth of castor bean callus cells was studied in vitro by Huang et al. (2017a), who were able to determine the distribution and the chemical form of Cu in the cells. Conversely, there is currently a lack of knowledge on the interaction of callus cell cultures (*Brassica napus*) with actinides and lanthanides with regard to their bioassociation and distribution, as well as their impact on cell growth and metabolism. Moreover, the speciation of actinides and lanthanides in callus cells and their cell compartments has yet to be fully investigated.

Accordingly, this study was designed to determine the tolerance of canola (*Brassica napus*) callus cells to U(VI) and Eu(III) at two different metal concentrations. The effects of both PTMs on cell growth and vitality, as well as on the total phenolic content of the cells, were studied. Furthermore, this investigation also focused on the speciation of bioassociated U(VI) and Eu(III) and their distribution in various fractions of *B. napus* cells, since *B. napus* is known to be able to accumulate PTMs in higher quantities than many other species (Laurette et al. 2012b).

**Materials and methods**

**Cell cultivation in the presence of Eu(III) and U(VI)**

*Brassica napus* callus cells were obtained from DSMZ (PC-1113, Braunschweig, Germany). The cells were cultivated in a 4-week growth cycle in the dark at room temperature on a solid modified Linsmaier and Skoog medium (medium R) containing 0.8% agar (Linsmaier and Skoog 1965). The callus cells were grown on a solid medium R with a reduced phosphate concentration of 6.25 × 10$^{-6}$ M (medium R$_{red}$, Tab. S1) supplemented with 20 or 200 μM UO$_2$(NO$_3$)$_2$ or 30 or 200 μM EuCl$_3$ (99.999%, Aldrich, Taufkirchen, Germany). The original phosphate concentration of the medium was reduced to minimize the precipitation of Eu(III) and U(VI) phosphate complexes.

Friable callus cells (400 mg) were transferred into Petri dishes (Roth, Karlsruhe, Germany) with the respective PTM-containing medium R$_{red}$. The Petri dishes were then sealed with Parafilm®M (Bems, Brain L’Alleud, Belgium) and stored in the dark at room temperature. Control samples lacking either Eu(III) and U(VI) were prepared under the same conditions. Eight independent experiments were performed with at least three, and at most ten, parallel samples used for controls with each heavy metal concentration. Cell growth was monitored every week. After about 6 weeks, the cells were collected from the solid medium, combined, and the resulting total weight was determined. In order to study the metal bioassociation by the cells as a function of the exposure time, samples from four experiments were collected weekly and analyzed.

**Determination of the Eu(III) and U(VI) bioassociation by the cells**

In order to determine the amount of bioassociated Eu(III) and U(VI), which represents the sum of the metal sorbed onto the
cells and taken up by the cells, about 100 mg fresh cells were weighed into 15-mL Greiner tubes (Greiner, Bio-one, Frickenhausen, Germany). A mixture of 2-mL concentrated HNO₃ (≥ 65%, p.a., Roth) and 1.5 mL 30% H₂O₂ (p.a., stabilized, Roth) was added; the cells were heated in a water bath at 80 °C for 7 h to digest the cells. Then, Milli-Q water was added to the solutions to reach 10 mL total volume, which were then analyzed for their U, Eu, Mg, and Ca content by inductively coupled plasma-mass spectrometry (ICP-MS; model ELAN 9000, PerkinElmer, Boston, MA). The results represent mean values and standard deviations of the mean.

Vitality measurements

The cell vitality was determined using an MTT assay (Mosmann 1983) as described in Sachs et al. (2017). This approach measures the activity of mitochondrial and cytosolic dehydrogenases, which reduces the yellow, water soluble 3-(4,5-dimethylthiazol-2-yl)-2,5-di-phenyl-tetrazolium bromide (MTT) to a blue, water-insoluble formazan product (Lindl and Gstraunthaler 2008).

After cell exposure to Eu(III) or U(VI), 50 mg of fresh cells were weighed into 1.5-mL reaction tubes (Greiner) followed by the addition of 1-mL phosphate-buffered saline solution without Ca²⁺ and Mg²⁺ (PBS; Biochrom, Berlin, Germany) and 200 μL MTT solution (5 mg/mL; Duchefa, Harlem, The Netherlands). Subsequently, the assay was performed as described in Sachs et al. (2017). The vitality of the Eu(III) and U(VI) exposed cells was determined as a percentage of the control samples according to Eq. (1).

\[
\text{Cell vitality (of control)} = \frac{\text{Absorbance of exposed cells}}{\text{Absorbance of control cells}} \times 100
\]

(1)

The results represent data from five independent experiments, each with two to three samples for control and each metal concentration.

Estimation of phenolic compounds

The total phenolic content of the callus cells after 39 to 43 days growth time in the absence or presence of Eu(III) or U(VI) was estimated based on Ainsworth and Gillespie (2007). About 100 mg of the fresh cells were added to 1.5-mL reaction tubes (Greiner Bio-one) and then immediately frozen in liquid nitrogen. Subsequently, 1 mL of ice-cold 95% (vol/vol) methanol (Roth) was added to the frozen cells after which the cells were homogenized by applying a plastic pestle. The samples were incubated in the dark for 48 h at room temperature. After incubation, the samples were centrifuged (13,000×g, 5 min, room temperature; centrifuge 5415R, Eppendorf, Hamburg, Germany) and the supernatants were separated. A total of 100 μL of each supernatant was pipetted into fresh 1.5-mL reaction tubes; 200 μL of 10% (vol/vol) Folin–Ciocalteu reagent (2 N; Merck, Darmstadt, Germany) was added and the samples were thoroughly mixed on a vortex mixer (Rex control, Heidolph, Schwabach, Germany). After 1–2 min, 800 μL 0.7 M Na₂CO₃ (p.a., Roth) solution was added to each tube. All samples were thoroughly mixed and subsequently incubated in the dark for 2 h at room temperature. Under the same conditions, both standard and blank solutions were prepared starting with 0.05–1.0 mM gallic acid (98%, Acros, Geel Belgium) stock solutions in 95% methanol and 95% methanol, respectively. After incubation, 8 × 100 μL sample, standard, or blank solutions were transferred into 96-well plates and the absorbance at 620 nm was measured in a microplate reader (Mithras LB940, Berthold, Bad Wildbad, Germany). A standard curve was calculated from the blank corrected gallic acid standards considering three phenol equivalents per gallic acid molecule. The total phenolics of the blank corrected samples were estimated using the regression equation from the gallic acid standard curve. The results represent mean values and standard deviations of the mean of the eight independent experiments with 25 individual samples for control and each metal concentration.

Statistical analyses

The statistical evaluation of selected experimental data (cell growth, cell vitality, phenolic compounds, and Ca(II) + Mg(II) cell contents) was performed by the two-tail Student’s t-test. The statistical analyses were done using the implemented functions in the “Analysis ToolPak” of Microsoft Excel 2010. The p value was used to discriminate between data groups showing significance (< 0.05) and those that were not. One asterisk (p value less than 0.05) denotes statistical significance and two asterisks is a measure for very significant events (p value less than 0.01). P values less than 0.5 can be interpreted by a tendency visible in the course of experimental data. The boxplots were prepared also with Microsoft Excel 2010.

Cell fractionation experiments

Cells that were grown under identical conditions were collected from the solid medium Rred and combined to a bulk sample. Approximately, 3 g of the cells was suspended in 5 mL ice-cold 0.154 M NaCl (p.a., Roth) for Eu(III) or 0.154 M NaNO₃ (99%, Sigma, Steinheim, Germany) for U(VI). The pH of the NaCl and NaNO₃ solutions was 5.8. These suspensions were then transferred to a glass homogenizer where the cells were homogenized. Cell fractions were separated by differential centrifugation of the cell homogenate (15 min at 1000×g; 15 min at 30,000×g; 60 min at 50,000×g) (Centrifuge 5804R, Eppendorf; Sorvall Evolution RC, Kendro,
Langenselbold, Germany). Pellet 1 consisted of heavy cell residues. Pellet 2 represents the lighter cell components (e.g., cell organelles). Pellet 3 contained the membrane-containing fraction and destroyed cell organelles. The last supernatant represents the soluble components (e.g., macromolecular organic matter and inorganic ions including soluble fractions from broken organelles) in the cytosol. The assignment of the fractions was based on the work of Carrier et al. (2003). The U(VI) and Eu(III) speciation in all three pellets and the cytosol fractions were investigated by TRLFS as described below.

In order to determine the Eu(III) and U(VI) content of the individual fractions, aliquots of Pellets 1 to 3 were digested with a mixture of HNO$_3$ and H$_2$O$_2$ as described above. The resulting solutions were analyzed together with the cytosol fraction that was acidified with HNO$_3$ for their U and Eu content by ICP-MS.

### Time-resolved laser-induced fluorescence spectroscopy measurements

#### Europium(III)-TRLFS

Eu(III) TRLFS studies were performed as described in Moll et al. (2009) and Moll et al. (2014). Resuspended cells and the cell fractions were measured, which were placed in 1-cm quartz glass cuvettes (Hellma Analytics, Mühlheim, Germany). For this phase of the investigation, cells or cell fractions were suspended in 0.154 M NaCl (pH 5.8) or measured directly at room temperature. Static emission spectra were recorded from 564 to 648 nm with the 1200 lines/mm grating, a dynamic step width was used to describe species with a long emission lifetime, as well as species with short emission lifetime. The following formula was used:

$$t_i = t_0 + F_1 + F_2 \cdot x$$

- $t_0$ initial delay, set to 1 $\mu$s
- $F_1$ factor 1, set to 5 $\mu$s
- $F_2$ factor 2, set to 1 $\mu$s
- $i$ number of spectrum
- $x$ number of previous spectrum

The abatement of the luminescence was investigated over 50 time points, resulting in 50 spectra. The spectra of individual samples were averaged, baseline and energy-corrected, and normalized. The spectra were normalized to the area of the $^7F_2$ band of Eu(III) with OriginPro 8.6.0G (OriginLab Corporation, USA). Luminescence emission lifetimes were determined with a non-linear fitting as exponential function (ExpDecay1, ExpDecay2) with the same software. The relative peak intensity ratio, $R_{E/M}$, which gives information about the ligand field of Eu(III) and the coordination environment, was determined by forming the ratio for the integral intensities of the $^7F_2$ to $^7F_1$ band, as presented in Eq. 3:

$$R_{E/M} = \left( \frac{\langle 5D_0 \rightarrow 7F_2 \rangle}{\langle 5D_0 \rightarrow 7F_1 \rangle} \right)$$

The intensities of the transitions ($^5D_0 \rightarrow ^7F_2$) and ($^5D_0 \rightarrow ^7F_1$) were calculated from the corresponding normalized peak areas. The number of coordinated water molecules was determined based on the equations of Kimura and colleagues (Kimura and Choppin 1994; Kimura et al. 1996; Kimura and Kato 1998), which is presented for europium in Eq. 4:

$$N_{H_2O} = 1.07 k_{exp}^{-0.62}$$

- $N_{H_2O}$ coordination number of water molecules
- $k_{exp}$ reciprocal luminescence emission lifetime (ms)

### Uranium(VI)-TRLFS

U(VI) TRLFS studies were performed as described in Geipel and Viehweger (2015) and Sachs et al. (2017). Fresh cells or cell fractions were suspended in 0.154 M NaNO$_3$ (pH 5.8) and deposited into quartz cuvettes (Hellma Analytics). Spectra were recorded at room temperature using a Peltier-controlled cuvette holder (Flash 300; Quantum Northwest, USA), which was set to 293 K. Depending on the uranium concentration, spectra were measured with 40 to 100 laser pulses per spectrum in the wavelength range between 450.4 and 727.0 nm at a resolution of 0.266 nm. The spectra were evaluated with the OriginPro 2015G software (OriginLab Corporation). To confirm the bioassociation of U(VI) with the cells, reference solutions of 20 or 200 $\mu$M U(VI) in 0.154 M NaNO$_3$ (pH 5.8) were produced and measured using a coupled Minilite I and Minilite II Nd:YAG laser system (Continuum Electro Optics Inc., Santa Clara, USA) with a repetition rate of 10 Hz. The excitation wavelength was set at 266 nm with pulse energies of about 0.3 mJ. The luminescence emission was focused into a spectrograph (iHR 550, Horiba Jobin Yvon GmbH, Munich, Germany) and detected using an intensified camera system (Horiba Jobin Yvon). Spectra were measured by averaging 100 laser pulses per spectrum in the wavelength range between 372.2 and 671.6 nm at a resolution of 0.461 nm.

### Results and discussions

#### Cultivation of B. napus cells in the presence of Eu(III) and U(VI): Cell growth, vitality, and bioassociation of both metals as a function of their concentration

Figure 1a illustrates the development of the callus cell samples in the presence of Eu(III) or U(VI) with increasing exposure time compared to control samples. Note that for all samples,
both an increase in the amount of cells and a darkening of the cells were visible with increasing growth time. However, after 6 weeks, the amount of cells grown in the presence of 200 μM Eu(III) or U(VI) appeared to be lower compared to the other samples. This reduction is reflected in the increase of the total cell weight, which is illustrated in Fig. 1b for each sample and the control. For each experiment, accumulated cells from all individual plates were collected and combined in one bulk sample at the end of the experiment. Data in Fig. 1c represent values from five independent experiments, each with 2–3 samples for control and each metal concentration. Significant differences to the untreated cells were calculated by Student’s t test and are indicated by * (p < 0.05) or ** (p < 0.01), respectively.

Fig. 1 B. napus callus cells grown in the presence of different Eu(III)/U(VI) concentrations after different exposure times (a). Box plots for the increase in fresh cell weight (b) and for the cell vitality of B. napus callus cells (c) grown on solid medium R red after an exposure time of 6 weeks. Data in Fig. 1b represent measurements from seven independent experiments. For each experiment, accumulated cells from all individual plates...
analysis showed only a slight tendency of the noticed Ca(II) effect. More independent experiments would be needed to verify the postulated Ca(II) effect.

Eu(III) is not only an analogue for trivalent actinides (e.g., Cm(III) and Am(III)), but it serves as an analogue for Mg(II) and Ca(II) as well. To date, research indicates that substituting Eu(III) for Ca(II)/Mg(II) does not appear to impact the normal physiological functions of plant tissues; however, the phyto-toxic effects of Eu(III) are unknown (Fellows et al. 2003). Gao et al. (2003) reported that Eu(III) can directly enter into plant cells through the Ca2+ ion channel and competes with Ca2+ for protein binding sites. U(VI) in the form of the UO22+ cation also resembles Ca2+ and Mg2+ and is able to form complexes with higher stabilities (Vandenhove et al. 2006). Additionally, it is known that U(VI) can replace Ca2+ and Mg2+, which can lead to structural changes in cell membranes, enzyme inactivation, and damage to RNA and DNA (Saenen et al. 2013). Our observation of a slightly increased Ca(II) uptake with no significant change in the Mg(II) content of the cells in the presence of 200 μM Eu(III) or U(VI) is comparable to the findings of Vanhoudt et al. (2010), who reported enhanced calcium uptake and almost unchanged magnesium concentration in the roots of Arabidopsis thaliana seedlings in the presence of uranium and cadmium. Küpper and Kochian (2010) reported that enhanced calcium uptake under cadmium stress in Thlaspi caerulescens could impede the replacement of calcium by cadmium in proteins as a defense mechanism against cadmium toxicity. More recently, Huang et al. (2017b) described the protective role of Ca against Cd-induced toxicity in plants. In a similar study, Cao et al. (2018) described how the homeostasis of Ca and Mg in Camellia sinensis after Cd treatment was affected at Cd concentrations between 1 and 15 mg/L, noting that the intracellular Ca content in leaves increased with increasing Cd stress, but was much less pronounced for Mg. As documented by Yang et al. (2015), treating horseradish roots with Tb(III) resulted in Tb(III) accumulation in both the extracellular and intracellular spaces of the roots, accompanied by increasing intracellular Ca content as well. These various studies confirm that plant cells do respond to PTM stress with an increased concentration of intracellular Ca.

**Estimation of extractable phenolic compounds**

Researchers have documented that plants may synthesize protective metal binding metabolites, e.g., phenolic compounds such as flavonoids or hydrocinnamic acid, in response to PTM stress (Wu et al. 2013). These metabolites can complex the metal ions, store them in the vacuole, or secrete them into the rhizosphere. It should also be noted that phenolic compounds, like flavonoids and lignin precursors, can scavenge harmful oxygen species arising from the unwanted effects of PTM stress (Sytar et al. 2013). In order to study the influence of Eu(III) and U(VI) on the phenolic substance pool of B. napus cells, we estimated the total phenolic content of the cells after PTM exposure by applying the Folin-Ciocalteu assay and then comparing results to our control samples. It is important to add that the results of this assay must be interpreted as an estimation, since other oxidizable substrates (i.e., in addition to phenolic compounds) can also react with the Folin-Ciocalteu reagent. In general, an increase in the phenolic content of the cells was detected within the studied time frame. Figure SI1 shows the mean values of the total phenolic content of the metal-exposed plant cells after 6 weeks of exposure to Eu(III) or U(VI) in comparison to analogous data for the control samples. Note that the mean values for the phenolic equivalents exhibited no clear trend based on the different exposure conditions. Moreover, the phenolic content of the cells exposed to 30 or 200 μM Eu(III) turned out to be quite similar to those of the control samples. In the case of U(VI), we detected a slight tendency toward increase in the total phenolic content for those cells exposed to 20 μM U(VI), whereas in the presence of 200 μM U(VI), this value was comparable to those of the controls. Similar behavior was also reported for the effects of Cd exposure on tea plant callus cultures from the roots and stem (Zagoskina et al. 2007). In the presence of 63 μM Cd, an increase in the phenolic content of the cells was observed in comparison to the control sample, whereas...
in the presence of 106 μM Cd, this value was close to the control.

However, in spite of the darkening of the cells (cf. Fig. 1a) that suggests an increase in the content of lignin and its precursor substances, we did not observe any significant effect of the presence of Eu(III) or U(VI) on the extractable phenolic content of the cells.

**Cell fractionation experiments**

To determine the speciation of Eu(III) and U(VI) in the different cell compartments of plant cells, cell fractionation experiments were performed. The fractionation protocol described in Geipel and Viehweger (2015) was modified in order to avoid anticipated difficulties in distinguishing the metal speciation in the buffer system and within the cell compartments. Specifically, we replaced the complex buffer solutions with the use of 0.154 M NaCl for Eu(III) and 0.154 M NaNO₃ for U(VI). The Eu(III)/U(VI) distribution in the different cell fractions is depicted in Fig. 3. As shown therein, most of the Eu(III) and U(VI) was bound onto heavy cell components, e.g., the cell wall fraction (Pellet 1), which can be explained by the function of the cell wall—and its high metal-absorption capacity—as the first barrier in preventing metals from entering the cellular environment. This finding is in agreement with the distribution of copper in castor bean callus cultures (Huang et al. 2017a). It should be noted, however, that a small amount of U(VI) (16 μg/g fresh cells at 200 μM U(VI)) and Eu(III) (5 μg/g fresh cells at 200 μM Eu(III)) was found within the cells in the cytosol fraction, thus indicating that PTM uptake into the cells had occurred (cf. Fig. 3).

**Spectroscopic analysis of Eu(III) and U(VI) speciation**

TRLFS is a non-invasive, selective, and highly sensitive method for detecting Eu(III) and U(VI) in the nM to μM concentration range, with resulting spectroscopic data used to conduct speciation analyses in order to determine the local environment of both metals (Geipel 2006; Binnemans 2015).

**Results of Eu(III)-TRLFS**

Following an exposure time of 6 weeks, we obtained luminescence spectra of Eu(III) bioassociated by B. napus cells as a function of the initial Eu(III) concentration, which are depicted in Fig. 4. Eu(III)-loaded plant cells were either measured directly or suspended in 0.154 M NaCl, with no differences in the luminescence spectra and lifetimes observed. Our results confirmed that the intensity of the hypersensitive 5D₀ → 7F₂ transition at about 616 nm increased strongly; the symmetry-forbidden 5D₀ → 7F₀ transition appeared at around 579 nm (Fig. 4a); and the luminescence decay changed to bi-exponential with prolonged lifetimes (Table 1, Fig. 4b). The luminescence spectrum of the Eu³⁺ aqua ion was shown to be characterized by emission bands at 585–600 nm (magnetic dipole transition 5D₀ → 7F₁) and 610–625 nm (hypersensitive transition 5D₀ → 7F₂). The intensity ratio according to Eq. (3) of 0.5 and the measured lifetime of 112 ± 5 μs corresponding to 9 water molecules in its first coordination sphere are in good agreement with the literature (e.g., Horrocks and Sudnick 1979; Kimura and Choppin 1994; Kim et al. 1994; Moulin et al. 1999; Planque et al. 2003; Heller et al. 2012; Barkleit et al. 2013).

Although the 7F₁ peak should not be influenced by complexation, for all our plant cell suspensions, we observed a slight decrease in intensity combined with a broadening of this transition (cf. Fig. 4a). The interaction of Eu(III) with B. napus cells was noted to be especially pronounced in the 7F₂ peak. The Eu(III) bioassociated by the plant cells appeared to be characterized by strongly enhanced Rₑ/M values of 4.3 ± 0.3 and 3.7 ± 0.6 for cells grown in the presence of 30 and 200 μM Eu(III), respectively (Table 1).

This finding indicates that the plant cells established a strong ligand field to Eu(III), with the resulting formation of strong species. We also note that bi-exponential luminescence decay was detected in plant cell samples, indicating the occurrence of two different Eu(III) coordination environments. The lifetime of the short-lived component varied between 120 (200 μM Eu(III)) and 150 μs (30 μM Eu(III)). According to Eq. (4), 8 to 6 coordinated water molecules should remain.
Assuming that Eu(III) maintains a nine-fold coordination, 1 to 3 binding sites will be filled up by functionalities of the plant cell envelope. However, the lifetime of the long-lived component varied between 460 (200 \( \mu \)M Eu(III)) and 750 \( \mu \)s (200 \( \mu \)M Eu(III)). According to Eq. (4), only 1 to 2 coordinated water molecules should remain, with 7 to 8 binding sites filled up by functionalities of the plant cell. This strong change in the hydration sphere of Eu(III) points to a bioassociation of Eu(III) with \( B. \) napus cells. Time-dependent luminescence measurements of Eu(III) (30 and 200 \( \mu \)M) in solid medium \( R_{red} \) prior to cell contact (see Fig. SI2 and Table 1) showed a different Eu(III) speciation found on the cells. Calculating the Eu(III), as well as the U(VI) speciation in the solid medium \( R_{red} \), was challenging due to missing stability constants with the individual medium components (see Table SI1).

The resulting luminescence emission spectra and the corresponding luminescence decays for the Eu(III) found in the different cell fractions are depicted in Fig. 4c and d. In comparing the spectral parameters of Eu(III) taken up by the plant cells with those in the cell fractions, we noted a decrease in \( R_{E/M} \) values. Also, we measured prolonged lifetimes in the sequence cells, Pellet 1, Pellet 2, and Pellet 3 independently of the initial Eu(III) concentration (Table 1). The lowest \( R_{E/M} \) values of 2.3 and 2.8 (200 \( \mu \)M and 30 \( \mu \)M Eu(III)) were detected within the cytosol of the inner part of the cell. In contrast to the other fractions, here the short-lived component clearly dominated with 80% of total luminescence decay. This finding points to another Eu(III) speciation in the cytosol than just in the outer areas of the cell fragments. Our results indicate that the sum spectra of Eu(III) bioassociated to \( B. \) napus cells was dominated by the influence of large cell fragments (e.g., Pellets 1 and 2).

Based on lifetime measurements, Ozaki et al. (2002) confirmed a relationship between experimentally obtained \( R_{E/M} \), the strength of the ligand field, and the geometrical structure around Eu(III). This previously reported empirical approach, which relies on the construction of coordination environment (CE) diagrams, was found to be effective for predicting the coordination environment of both the hydrated and complexed Eu(III) in solutions, as well as that of adsorbed Eu(III) on both the ion-exchange resins and on mammalian cells and microorganisms (Sachs et al. 2015; Moll et al. 2014). Our interpretations of the CE diagrams that we developed are based on the earlier work of Ozaki et al. (2002), as follows. In solutions in which Eu(III) interacts with ligands other than water in an outer-spherical manner, \( R_{E/M} \) increases with increasing interaction, whereas \( \Delta N_{H2O} \) remains small (between 0 and 3), resulting in scattered data from the lower left to the vicinity of the origin at the lower right corner.
right area of the CE diagram. In the case of predominant inner-
sphere coordination, \( R_{E/M} - \Delta N_{H2O} \) data is scattered in the
upper-left area of diagram (Fig. 5). Moreover, luminescence
measurements as a function of the delay time indicate that the
\( R_{E/M} \) value of the sum spectrum is influenced principally by
the long-lived component. Therefore, the CE diagram present-
ed in Fig. 5 is based on the lifetimes of the long-lived
components.

First of all, strong inner-sphere complexes were formed in
all samples, which are reflected in the scattered data in the
upper part of the diagram (cf. Fig. 5). The Eu(III) coordination
environment in the cytosol showed similarities with (a) Eu(III)
bound to the bacterial phosphate groups of the cell envelope of
Sporomusa sp. and bacterial lipopolysaccharide (Moll et al.
2014; Bader et al. 2019), (b) Eu(III) complexed by the strong
chelate-ligand EDTA, and (c) Eu(III) complexed by carboxyl
groups of salicylic acid (Barkleit et al. 2013). In this first
approximation, we confirmed the interaction of Eu(III) with
organic phosphate and carboxyl groups in a chelate-manner in
the cytosol at the level of \( [\text{Eu(III)}]_0 \). 200 \( \mu \)M. Additionally,
Eu(III) on/in \( B. \) napus cells and in the cytosol at the level
30 \( \mu \)M \( [\text{Eu(III)}]_0 \) was found to be characterized by an even
stronger inner-sphere character.

On closer inspection of the luminescence data of \( B. \) napus
cells exposed to 30 or 200 \( \mu \)M Eu(III), slight differences in the
Eu(III) speciation can be deduced. Cells exposed to 30 \( \mu \)M
Eu(III) depicted a higher \( R_{E/M} \)-value and prolonged
luminescence lifetimes compared with cells exposed to
200 \( \mu \)M Eu(III). Hence, at 30 \( \mu \)M Eu(III), plant cells
established a more intense ligand field to Eu(III). Here both
Eu(III) species contained less coordinated water molecules, 1
or 6, respectively. Consequently, more functional groups pro-
vided by the cells are involved in the respective Eu(III) species
compared with the two Eu(III)-species formed in the presence
of 200 \( \mu \)M Eu(III). The more intense Eu(III) interaction to
plant cells exposed to 30 \( \mu \)M Eu(III) is also depicted in the
CE diagram (cf. Fig. 5).

Drake et al. (1997) assessed \( ^7F_0 \rightarrow ^5D_0 \) transition excitation
spectra to examine the binding sites on native \( Datura \) innoxia
cell wall fragments, with four unique binding sites reported to
be involved in metal ion uptake. In particular, the researchers
reported that higher-affinity sites tended to involve carboxyl-
ates. The native \( Datura \) innoxia cell wall fragments treated
with 300 to 3000 \( \mu \)M Eu(III) at pH 5 also exhibited a bi-
exponential luminescence decay with the two lifetimes of
263 and 630 \( \mu \)s (Table 1). The authors concluded that the
shorter lifetime is consistent with a 1:1 carboxylate complex-
ation, whereas the longer lifetime indicated that a second and
third carboxylate are bound. By comparing these lifetimes
with our results, which includes the Biorex values (Drake
et al. 1997), an involvement of \( B. \) napus cell carboxylates
seems possible. However, we cannot discount the potential
contribution of other functionalities (e.g., phosphate moieties).
Most likely, \( B. \) napus cells provide multiple-binding

| Sample composition            | \( F_2/F_1 \) ratio (\( R_{E/M} \)) | Lifetime 1/\( \mu \)s | \( N_1 \) H\( _2 \)O | Lifetime 2/\( \mu \)s | \( N_2 \) H\( _2 \)O |
|------------------------------|------------------------------------|----------------------|-----------------|----------------------|------------------|
| Eu\( ^{3+} \) in 0.154 M NaCl pH 5.8 (30, 200 \( \mu \)M) | 0.52 ± 0.03 | 112 ± 5 | 8.9 ± 0.2 |
| 30 \( \mu \)M Eu(III) | | | |
| 30 \( \mu \)M Eu(III) in solid medium \( R_{red} \) | 3.0 ± 0.1 | 133 ± 5 (92%) | 7.4 ± 0.3 | 364 ± 31 (8%) | 2.3 ± 0.2 |
| Eu(III) on callus cells | 4.3 ± 0.3 | 150 ± 6 (69%) | 6.5 ± 0.3 | 750 ± 24 (31%) | 0.8 ± 0.03 |
| Pellet 1 | 4.1 ± 0.3 | 164 ± 9 (61%) | 5.9 ± 0.3 | 783 ± 36 (39%) | 0.7 ± 0.03 |
| Pellet 2 | 4.1 ± 0.2 | 173 ± 10 (56%) | 5.6 ± 0.3 | 800 ± 35 (44%) | 0.7 ± 0.03 |
| Pellet 3 | 3.7 ± 0.3 | 212 ± 8 (57%) | 4.4 ± 0.2 | 853 ± 16 (43%) | 0.6 ± 0.02 |
| Cytosol | 2.8 ± 0.5 | 132 ± 5 (79%) | 7.5 ± 0.3 | 594 ± 13 (21%) | 1.2 ± 0.03 |
| 200 \( \mu \)M Eu(III) | | | |
| 200 \( \mu \)M Eu(III) in solid medium \( R_{red} \) | 2.7 ± 0.1 | 125 ± 5 (84%) | 7.9 ± 0.3 | 285 ± 16 (16%) | 3.1 ± 0.2 |
| Eu(III) on callus cells | 3.7 ± 0.6 | 121 ± 5 (54%) | 8.2 ± 0.3 | 460 ± 8 (46%) | 1.7 ± 0.04 |
| Pellet 1 | 3.1 ± 0.2 | 123 ± 7 (48%) | 8.1 ± 0.5 | 470 ± 10 (52%) | 1.7 ± 0.05 |
| Pellet 2 | 2.9 ± 0.1 | 139 ± 7 (46%) | 7.1 ± 0.4 | 516 ± 10 (54%) | 1.5 ± 0.03 |
| Pellet 3 | 2.8 ± 0.5 | 170 ± 9 (55%) | 5.7 ± 0.3 | 553 ± 16 (45%) | 1.3 ± 0.04 |
| Cytosol | 2.3 ± 0.02 | 144 ± 4 (80%) | 6.8 ± 0.2 | 424 ± 26 (20%) | 1.9 ± 0.1 |
| Native \( Datura \) innoxia cell wall fragments pH 5 (Drake et al. 1997) | 263 ± 14 | 3.4 ± 0.2 \( ^a \) | 630 ± 19 | 1.1 ± 0.03 \( ^a \) |
| Biorex (carboxylate) pH 5 (Drake et al. 1997) | 282 ± 5 | 3.2 ± 0.2 \( ^a \) | 620 ± 17 | 1.1 ± 0.03 \( ^a \) |
| Eu\( ^{3+} \)- Oat (\( Avena \) sativa) roots (Fellows et al. 2003) | 345 | 2.8 ± 0.5 |

\( N \), number of coordinated water molecules
\( ^a \) Calculated in this study
environments for Eu(III), while some binding sites showed relatively poor luminescent properties. The observed differences, especially for the shorter lifetime, can be attributed to additional processes taking place within living cells in comparison to dead (i.e., inert) biomass. Fellows et al. (2003) conducted an in situ investigation of Eu(III) uptake in oat (Avena sativa) roots by TRLFS, indicating that Eu(III) uptake was highest within undifferentiated root cells. B. napus callus cells are also undifferentiated plant cells, which have the propensity to bioassociate a considerable amount of Eu(III). The spectral changes that Fellows et al. (2003) observed when Eu(III) was complexed by cellular components of the oat root indicated the involvement of carboxylic and amino carboxylic functionalities. In summary, the authors linked their data findings to a strong inner-sphere mononuclear Eu(III) complex inside the root with a luminescence lifetime of 345 μs (Table 1). Similarly, our results for Eu(III) in the B. napus callus cell system also demonstrated the formation of strong inner-sphere Eu(III) complexes—although the luminescence-decay behavior varied (Table 1). Again, the varying luminescence decay behavior with different lifetime points to multiple-binding environments in living B. napus cells.

Results of U(VI)-TRLFS

Figure 6 provides static luminescence spectra for the B. napus cells cultivated in the presence of 20 or 200 μM U(VI) compared to the luminescence spectra of U(VI) in 0.154 M NaNO₃ (pH 5.8). In contrast to our Eu(III) assays, dynamic quenching processes significantly decreased the luminescence lifetimes of the U(VI) species due to the presence of high amounts of both organic substances and Fe₃⁺ and Cl⁻, thereby hampering the ability to compare our findings with model substances as already discussed in Sachs et al. (2017). As a consequence, the measured spectra were not analyzed with regard to their lifetimes. In addition, due to the complex U(VI) luminescence quench processes taking place in the presence of organic substances, TRLFS measurements of the U(VI)-containing solid

**Fig. 5** CE diagram of Eu(III) bound to B. napus callus cells as Eu(III) species with the long lifetime (y-axis: number of water molecules in the inner-sphere: ΔN₁₂₂₀ = 9-N₁₂₂₀ and x-axis: strength of ligand field RₑM = relative peak intensity ratio according to Eq. (3)). EDTA: ethylenediaminetetraacetic acid, LPS: lipopolysaccharide. References: (a) Moll et al. 2014, (b) Ozaki et al. 2005, (c) Ozaki et al. 2006, (d) Heller et al. 2012, (e) Barkleit et al. 2013, (f) Bader et al. 2019

**Fig. 6** TRLFS spectra of B. napus cells cultivated in the presence of 20 μM U(VI) (a: 28 μg U/g fresh cells) and 200 μM U(VI) (b: 257 μg U/g fresh cells) after 39 days of exposure compared to TRLFS spectra of 20 or 200 μM U(VI) in 0.154 M NaNO₃ at pH 5.8
medium $R_{\text{red}}$ were not successful—again, in contrast to our Eu(III) trials. To compare the spectra of the individual samples and to identify dominant U(VI) species, spectra were analyzed by peak deconvolution using the peak-fitting module of OriginPro 2015G and compared to literature data for selected biological systems and U(VI) reference compounds (Table SI2).

Due to the low U(VI) concentration, the spectra of the cells that were cultivated in the presence of 20 μM U(VI) appeared to be dominated by the unspecific self-luminescence of the plant cells. Depending on the amount of bound U(VI), peaks were detected that point to the occurrence of bioassociated U(VI) (Fig. 6a). In contrast, spectra associated with cells grown in the presence of 200 μM U(VI) showed characteristic peaks, clearly indicating the occurrence of bioassociated U(VI) (Fig. 6b). An analysis of the spectra indicates a significant bathochromic shift of the emission bands of the plant cell species when compared to the spectra of the reference solutions of U(VI) in 0.154 M NaNO$_3$, which were dominated by the (UO$_2$)$_3$(OH)$_5$$^+$ species (cf. Fig. 6, Table SI2). This result points to the biologically induced binding of U(VI)—either extracellular on the cell surface or intracellular. A comparison of the main peak positions with those of U(VI) reference compounds indicates the dominant binding of U(VI) by organic and/or inorganic phosphate groups of the plant biomass (cf. Tab. S12). However, given the current state of knowledge, we were unable to distinguish between the coordination of the organic and inorganic phosphate groups of the cells. In addition, the dominant binding of U(VI) by carboxyl functionalities does not seem to be very important based on the fact that U(VI) carboxylate reference compounds demonstrated less pronounced bathochromic shifts of their emission peaks (cf. Tab. S12). Nonetheless, based on the data we obtained, we cannot exclude the potential, but minor, contribution of carboxylic compounds to the binding of U(VI) on the cell surface or within the cells. As expected, the main peak positions of the spectra of U(VI) bioassociated to $B$. napus cells are similar to those of U(VI) bioassociated with lupine roots from soil culture (Günther et al. 2003), green algae Chlorella vulgaris (Günther et al. 2008), and fungi Schizophyllum commune 12-43 (Günther et al. 2014) (cf. Tab. S12). Important for the current study, the authors of each of these prior spectroscopic investigations confirmed the predominant binding of U(VI) to inorganic and/or organic phosphate groups of the biomass. In addition, our TRLFS findings are supported by the EXAFS results of Laurette et al. (2012b), who suggested the complexation of U with intracellular inorganic and organic phosphate residues in $B$. napus and sunflower roots.

Figure 7 shows the normalized TRLFS spectra of cell fractions suspended in 0.154 M NaNO$_3$, which were obtained by fractionation of cells cultivated in the presence of 200 μM U(VI). In agreement with our ICP-MS results (cf. Fig. 3b), U(VI) was detected in all cell fractions, thereby indicating the extra- and/or intracellular bioassociation of U(VI). Note that the spectra of all fractions are similar in their main peak positions and intensity ratios, indicating a similar binding of U(VI). Differences in the signal-to-noise ratios can be attributed to the varying uranium content in the fractions, as well as the different sample amounts available for measurement. Note also that the spectra of the individual cell fractions are comparable to those of the whole cells. Therefore, the similarity of the spectrum for the whole cells when compared with that of Pellet 1, as well as the fact that most of the U(VI) was found in Pellet 1 (Fig. 3), confirms the dominant binding of U(VI) to heavy cell compartments (e.g., the cell wall), which represents an effective protective mechanism of the metabolically active cell compartments against the absorbance of potentially toxic metals. This result agrees with the findings of Laurette et al. (2012a), who described the predominant association of U with the insoluble structures of sunflower plants, e.g., cell walls.

In a related study, El Hayek et al. (2018) employed scanning transmission electron microscopy/energy-dispersive X-ray spectrometry to confirm the predominant binding of uranium in the cell walls of Brassica juncea. Nevertheless, it must be noted that for our investigation, the detection of U(VI) in the cytosol fraction indicates that a certain amount of U(VI) was taken up by the cells. This observation is in agreement with Geipel and Viehweger (2015), who studied the speciation of uranium in compartments of living $B$. napus cells in suspension cell cultures. Since different protocols for cell fractionation were used, however, the spectra reported by Geipel and Viehweger are not directly comparable to the spectra in the present work. While the positions of the main emission peaks are similar, the intensity ratios of the peaks differ slightly.

![Normalized luminescence spectra of $B$. napus cell fractions of cells that were cultivated in the presence of 200 μM U(VI)](image-url)
Conclusions

The results of *B. napus* callus cells grown in vitro under Eu(III) or U(VI) stress confirm that *B. napus* cells have a strong capacity to bioassociate both PTMs under the given experimental conditions. Most of the Eu(III) and U(VI) was bound on the cell wall fraction, which could represent the principal mechanism for Eu(III)/U(VI) enrichment. More likely, however, this finding points to the effective protective mechanisms of metabolically active cells against the threat of potentially toxic metals. We also note that, especially under high Eu(III)/U(VI) stress, both metals were found in the cytosol fraction, which does indicate the uptake of Eu(III)/U(VI) into the cells. High Eu(III)/U(VI) stress also showed the slight tendency that the homeostasis of Ca(II) in *B. napus* callus is affected. Moreover, this study confirmed that cell growth was reduced in combination with a decrease in cell vitality. The total cellular phenolic content, which could have increased due to PTM stress, was similar for cells that were exposed to 30 or 200 μM Eu(III), as well as for our controls. Only a slight tendency for a slightly higher phenolic content was found for cells grown in the presence of 20 μM U(VI), whereas at 200 μM U(VI), this value was lower to those of the controls.

Despite the low intensity of the symmetry-forbidden $^7F_0$ peak in the luminescence emission spectra of Eu(III) bound to *B. napus* cells, the appearance points to the formation of Eu(III) complexes. The occurrence of a bi-exponential luminescence decay confirmed the existence of two Eu(III) coordination environments. The strong intensity of the $^7F_2$ peak as a measure of changes in the Eu(III) speciation, coupled with the resulting high intensity ratio $R_{E/M}$, indicates the formation of strong Eu(III) complexes. Further analysis of the Eu(III) coordination environment revealed strong inner-sphere Eu(III) species, possibly with organic phosphate and carboxyl groups provided by the *B. napus* cells. In conclusion, *B. napus* cells provide multiple-binding environments for Eu$^{3+}$. Due to the bioassociation of U(VI) onto *B. napus* cells, we observed a significant bathochromic shift of the U(VI) emission bands of the plant cell species compared to the spectra of the reference solution (U(VI) in 0.154 M NaNO$_3$). Further, by comparing our findings with model compounds and other biological systems, there is a clear argument to be made for the predominant binding of U(VI) by organic and/or inorganic phosphate groups of the plant biomass. Our TRLFS-based speciation analysis confirmed this biochemical assessment in terms of the predominant binding of both Eu(III) and U(VI) on heavy cell compartments, such as the cell wall. In conclusion, *B. napus* cells are able to accumulate and tolerate potentially toxic metals like Eu(III) and U(VI). As a result, these metals have the potential to enter the food chain and may become a severe health risk for humans. We hypothesize that the PTM tolerance of these cells is likely due to several mechanisms—but most notably the strong binding of the metals within the cell walls that protects the cell compartments against heavy metal toxicity. The results of this study were obtained through combination of biological, biochemical, and spectroscopic methods. We hope that this integrative approach will contribute to an enhanced understanding of the interaction processes between actinides/lanthanides and plants at the molecular level, which is important for modeling the transfer of these elements in the environment.

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