Atmospheric hydrogen peroxide and organic hydroperoxides during PRIDE-PRD’06, China: their concentration, formation mechanism and contribution to secondary aerosols

W. Hua1, Z. M. Chen1, C. Y. Jie1, Y. Kondo2, A. Hofzumahaus3, N. Takegawa2, C. C. Chang4, K. D. Lu1,3, Y. Miyazaki5, K. Kita6, H. L. Wang1, Y. H. Zhang1, and M. Hu1

1The State Key Laboratory of Environmental Simulation and Pollution Control, College of Environmental Sciences and Engineering, Peking University, Beijing 100871, China
2Research Center for Advanced Science and Technology, Univ. of Tokyo, 4-6-1 Komaba, Meguro, Tokyo 153-8904, Japan
3Institut für Chemie und Dynamik der Geosphäre II: Troposphäre, Forschungszentrum Jülich, 52425 Jülich, Germany
4Research Center of Environment Changes, Academia Sinica, Nankang, Taipei 115, Taiwan
5Institute of Low Temperature Science, Hokkaido University, Kita-19, Nishi-8, Kita-ku, Sapporo 060-0819, Japan
6Faculty of Science, Ibaraki University, 2-1-1 Bunkyo, Mito, Ibaraki, 310-8512, Japan

Received: 25 April 2008 – Published in Atmos. Chem. Phys. Discuss.: 3 June 2008
Revised: 1 September 2008 – Accepted: 23 September 2008 – Published: 24 November 2008

Abstract. Atmospheric hydrogen peroxide (H2O2) and organic hydroperoxides were measured from 18 to 30 July in 2006 during the PRIDE-PRD’06 campaign at Backgarden, a rural site located 48 km north of Guangzhou, a mega-city in southern China. A ground-based instrument was used as a scrubbing coil collector to sample ambient air, followed by on-site analysis by high-performance liquid chromatography (HPLC) coupled with post-column derivatization and fluorescence detection. The H2O2 mixing ratio over the 13 days ranged from below the detection limit to a maximum of 4.6 ppbv, with a mean (and standard deviation) of (1.26±1.24) ppbv during the daytime (08:00–20:00 LT). Methyl hydroperoxide (MHP), with a maximum of 0.8 ppbv and a mean (and standard deviation) of (0.28±0.10) ppbv during the daytime, was the dominant organic hydroperoxide. Other organic peroxides, including bis-hydroxymethyl hydroperoxide (BHMP), peroxyacetic acid (PAA), hydroxymethyl hydroperoxide (HMHP), 1-hydroxy-ethyl hydroperoxide (1-HEHP) and ethyl hydroperoxide (EHP), were detected occasionally. The concentration of H2O2 exhibited a pronounced diurnal variation on sunny days, with a peak mixing ratio in the afternoon (12:00–18:00 LT), but lacked an explicit diurnal cycle on cloudy days. Sometimes a second peak mixing ratio of H2O2 was observed during the evening, suggesting that H2O2 was produced by the ozonolysis of alkenes. The diurnal variation profile of MHP was, in general, consistent with that of H2O2. The estimation indicated that in the morning the H2O2 detected was formed mostly through local photochemical activity, with the rest probably attributable to vertical transport. It is notable that relatively high levels of H2O2 and MHP were found in polluted air. The unexpectedly high level of HO2 radicals detected in this region can account for the production of hydroperoxides, while the moderate level of NOx suppressed the formation of hydroperoxides. High concentrations of hydroperoxides were detected in samples of rainwater collected in a heavy shower on 25 July when a typhoon passed through, indicating that a considerable mixing ratio of hydroperoxides, particularly MHP, resided above the boundary layer, which might be transported on a regional scale and further influence the redistribution of HOx and ROx radicals. It was found that hydroperoxides, in particular H2O2, play an important role in the formation of secondary sulfate in the aerosol phase, where the heterogeneous reaction might contribute substantially. A negative correlation between hydroperoxides and water-soluble organic compounds (WSOC), a considerable fraction of the secondary organic aerosol (SOA), was observed, possibly providing field evidence for the importance of hydroperoxides in the formation of SOA found in previous laboratory studies. We suggest that hydroperoxides act as an important link between sulfate and organic aerosols, which needs further study and should be considered in current atmospheric models.

Correspondence to: Z. M. Chen (zmchen@pku.edu.cn)

Published by Copernicus Publications on behalf of the European Geosciences Union.
1 Introduction

A series of hydroperoxides, including hydrogen peroxide \( \text{H}_2\text{O}_2 \) and organic hydroperoxides (ROOH), such as methylhydroperoxide (MHP, CH\(_3\)OOH), hydroxymethyl hydroperoxide (HMHP, HOCH\(_2\)OOH), 1-hydroxy-ethyl hydroperoxide (1-HEHP, CH\(_3\)CH(OH)OOH), peroxyacetic acid (PAA, CH\(_3\)C(O)OOH) and ethylhydroperoxide (EHP, CH\(_3\)CH\(_2\)OOH), have been measured in the atmosphere since the measurement of organic hydroperoxides was pioneered in the 1980s by Hellpointner and Gäb (1989). These reactive species play significant roles in atmospheric processes, such as acid precipitation, cycling of \( \text{H}_2\text{O}_2 \) radicals, and formation of secondary organic aerosol (SOA). \( \text{H}_2\text{O}_2 \) is considered to be the most important oxidant for the conversion of S (IV) to sulfuric acid and secondary sulfate in cloud, fog and rain water at pH < 5, thus contributing significantly to the acidification of clouds and rain (Penkett et al., 1979; Calvert et al., 1985; Fung et al., 1991; Pena et al., 2001). Organic peroxides such as MHP, HMHP, and PAA are able to oxidize \( \text{SO}_2 \), but only when \( \text{H}_2\text{O}_2 \) is limited (Lind et al., 1987; Zhou and Lee, 1992). For instance, HMHP is much more soluble than \( \text{H}_2\text{O}_2 \) (\( \text{H}_{2\text{MHP}}=7.7 \times 10^4 \text{M atm}^{-1} \) and \( \text{H}_{2\text{HMHP}}=1.7 \times 10^6 \text{M atm}^{-1} \) at 298 K, Sander et al., 2003) and can decompose rapidly into \( \text{H}_2\text{O}_2 \) and HCHO in the aqueous phase when pH > 5.0 (O’Sullivan et al., 1996; Chen et al., 2008). In addition, \( \text{H}_2\text{O}_2 \) and MHP can serve as temporary reservoirs of odd-hydrogen radicals (OH, \( \text{HO}_2 \), \( \text{CH}_3\text{O}_2 \)) in the troposphere, because their photolysis and other reactions will lead to the regeneration of OH radicals, and are intimately involved in the production of odd-oxygen (e.g. O, \( \text{O}_3 \)) (Madronich and Calvert, 1990; Lightfoot et al., 1992; Reeves and Penkett, 2003). For example, MHP, which is the most abundant organic hydroperoxide in the atmosphere, and has an atmospheric lifetime of 2–3 days and a low level of solubility in water (Cohan et al., 1999; Wang and Chen, 2006), can be transported to the upper troposphere at a regional scale without scavenging under deep convection conditions. As a result, this transportation probably leads to the redistribution of OH radicals in different regions and different altitudes (Jaeglé et al., 1997; Wennberg et al., 1998; Cohan et al., 1999; Mari et al., 2000; Ravetta et al., 2001), and \( \text{H}_2\text{O}_2 \) and organic hydroperoxides can be used as indicators of the oxidizing capacity of the troposphere (Thompson, 1992). Tropospheric aerosols play an important role in the Earth’s atmosphere and in the climate system. Aerosols scatter and absorb solar radiation (direct effect) (Andreae and Crutzen, 1997), change cloud characteristics in many ways (indirect effect) (e.g. Navakov and Penner, 1993; Lohmann and Feichter, 2005), and facilitate heterogeneous and multi-phase chemistry (Ravishankara, 1997). Increasing attention is being paid to the organic matter that represents a substantial fraction of tropospheric aerosols (Andreae and Crutzen, 1997). Recently, several laboratory studies have revealed that secondary organic aerosol (SOA) can be formed from isoprene and its gas-phase oxidation products through acid-catalyzed heterogeneous oxidation with hydrogen peroxide, a remarkably close analogy with atmospheric secondary sulfate formation (Claeys et al., 2004; Böge et al., 2006; Kroll et al., 2006).

No significant direct emission of \( \text{H}_2\text{O}_2 \) or organic hydroperoxides from natural or anthropogenic sources has been found, and it is believed that the majority of the \( \text{H}_2\text{O}_2 \) and ROOH in the gas phase are formed via the bimolecular and termolecular recombination of peroxy (HO\(_2\) and RO\(_2\)) radicals during the daytime. The only known mechanism for the formation of peroxides in the absence of light is the ozonolysis reaction of alkenes (Gäb et al., 1985; Becker et al., 1990, 1993; Valverde-Canossa et al., 2004), which is discussed in detail in Sect. 3.4. This reaction is the main source of the 1-hydroxyalkylhydroperoxides (1-HAHP) and a source of OH radicals (Atkinson and Aschmann, 1993; Paulson and Orlando, 1996).

Formation of HO\(_2\) radicals is predominantly through the photo-oxidation of carbon monoxide (CO) and volatile organic compounds (VOC) by the OH radical (described in detail by Lightfoot et al., 1992). The second significant part of HO\(_2\) is formed during the degradation of HCHO and other aldehydes by photolysis or by reaction with OH radicals (Buffalini et al., 1972; Su et al., 1979). Furthermore, the ozonolysis of alkenes, the decomposition of peroxy acetyl nitrate (PAN), and the photodegradation of aromatic hydrocarbons will provide a source of HO\(_2\) (Finlayson-Pitts and Pitts, 1986; Seuwen and Warneck, 1995). Alkylperoxy radicals (RO\(_2\)) are produced by the reaction of OH radicals with alkanes, e.g. \( \text{CH}_4 \), in the presence of oxygen, and by the decomposition of alkyl-substituted excited Criegee biradicals (Atkinson, 1994; Hatakeyama and Akimoto, 1994; Gäb et al., 1995).

The sinks for gaseous \( \text{H}_2\text{O}_2 \) and organic peroxides can be classified according to different processes, including washout through fog droplets and adsorption on water-covered aerosols or other wet surfaces; dry deposition; photolysis; and reaction with OH radicals. Although the importance of the individual processes might differ with regard to the water solubility of the organic peroxides (Gunz and Hoffmann, 1990; Watkins et al., 1995a, b), the washout and adsorption processes on wet surfaces are expected to be dominant.

Field, laboratory and modeling studies have all indicated that the generation and behavior of gas-phase \( \text{H}_2\text{O}_2 \) and organic ROOH are affected by the levels of chemical components such as NO\(_x\), CO, \( \text{CH}_4 \), and VOC. Additionally, meteorological parameters, including solar radiation, relative humidity, temperature, and pressure are of great importance in controlling the production and the loss of hydroperoxides (Lee et al., 2000).

Over the past two decades, the distribution and roles of \( \text{H}_2\text{O}_2 \) and CH\(_3\)OOH in the atmosphere have been investigated by various methods on land, onboard ship, and aboard
aircraft (Hellpointner and Gäb, 1989; Hewitt and Kok, 1991; Das and Aneja, 1994; Fels and Junkermann, 1994; Watkins et al., 1995a, b; Staffelbach et al., 1996; Heikes et al., 1996; Jackson and Hewitt, 1996; Sauer et al., 1997, 2001; Lee et al., 1993, 1995, 1997, 1998, 2000, 2008; Morgan and Jackson, 2002; Grossmann et al., 2003; François et al., 2005; Walker et al., 2006; Kim et al., 2007). Although numerous field measurements of H$_2$O$_2$ and organic peroxydes have been made, most of them were done at 25°–55° N, including North America, Brazil, Europe, Greenland, South Africa, and the Atlantic and the northwestern and central tropical Pacific (Lee et al., 2000). To our knowledge, data for hydroperoxides on land are not available for the East Asia low latitude region, where the atmospheric chemistry may be significantly distinguished from other regions on the earth. Accompanying rapid industrialization, East Asia has increasing amounts of O$_3$ precursor trace gases (carbon monoxide, nitrogen oxides, and hydrocarbons) released by industrial, agricultural and population growth. The Pearl River Delta (PRD) region, extending from the Hong Kong metropolitan area to the northwest, has been the most economically dynamic region of mainland China over the last two decades. The high levels of NO$_x$, SO$_2$, ozone and PM$_{2.5}$ observed in the PRD region over the past decade are believed to be associated with the rapid economic development (Zhang et al., 1998; Wang et al., 2003; Li et al., 2005). As intermediate photochemical byproducts, hydroperoxides can be used to test predictions by photochemical models by comparison with observed data (Jacob et al., 1996). Therefore, field studies of peroxydes are needed urgently to provide valuable data for investigating the photochemical mechanisms in this region and to be included in photochemical models.

We present a novel dataset for speciated hydroperoxides measured at a rural site in PRD that has high mixing ratios of VOC and CO. The objectives of this study were to investigate the impact of chemical and physical processes on the mixing ratio of H$_2$O$_2$ and organic peroxydes, to provide new field evidence of the existence of high mixing ratios of hydroperoxides above the planetary boundary layer (PBL), to examine the contribution of hydroperoxides to the formation of secondary sulfate and SOA, and ultimately to assess the value of hydroperoxide measurements for better understanding the mechanisms of secondary photochemical pollutions and to aid the development of more robust models.

2 Experimental

2.1 Measurement site

The measurements were carried out at Backgarden (23.548° N 113.066° E), a super-site for the PRIDE-PRD’06 (Program of Regional Integrated Experiments of Pearl River Delta Region) Air Quality Monitoring Campaign that took place from 3 to 30 July 2006. Backgarden is a rural site, surrounded by 20 km$^2$ of forest and 2.7 km$^2$ of lake, located north of the central PRD and about 60 km northwest of the mega-city Guangzhou in southeast China, which is the capital city of Guangdong Province. The increase in industry and population impacts the air quality of Guangzhou. Guangzhou is undergoing a complex air pollution composition due to the mixture of coal burning and vehicle emissions. Backgarden is a much less populated area at the outskirts of the densely populated center of the PRD and can be treated as a regional background site. Backgarden does not have significant local vehicle emission, while the biomass burning in the afternoon and cable burning in the evening might be a source of local emission. The Backgarden site experiences a typical sub-tropical climate and is usually influenced by the monsoon circulation in the July. During the period of hydroperoxides measurement, the observation site was mainly influenced by the western pacific subtropical high and typhoon. When controlled by the western pacific subtropical high pressure from 19 to 23 July, the days were sunny, and the dominant wind direction at Backgarden was southerly. When influenced by typhoon from 24 to 25 July, northerly winds prevailed at the observation site. During the last days of the campaign from 26 to 30 July, the local weather conditions were cloudy and rainy and the sampled air masses came mainly from the south/southeast. The average meteorological values (arithmetic mean ± standard deviation) for the campaign were: 29.5±3.4°C ambient temperature, 76.2±14.4% ambient relative humidity, 100.1±4 hPa ambient pressure, and 1.9±1.2 m/s local wind speed. The details of meteorological values are shown in Fig. 4 in Sect. 3.1. The primary pollutants including NO$_x$, CO, and SO$_2$ influenced the hydroperoxides level significantly during the campaign. Their concentrations kept relatively low in the daytime, but were elevated at night, particularly under southeasterly from 20 to 22 July. In addition to the measurement of peroxydes, all major trace gases (NO$_x$, NO$_y$, PAN, SO$_2$, CO, O$_3$, biogenic/anthropogenic VOC, etc.), aerosols (mass concentration, number concentration, chemical compositions), free radicals (OH, HO$_2$, RO$_2$) and meteorological parameters (temperature, wind direction, wind speed and relative humidity, pressure, rainfall) were monitored at this site by a number of groups.

2.2 Measurement method for hydroperoxides

2.2.1 Measurement method

A three-story hotel building, which located next to a 2.7 km$^2$ reservoir in a rural resort surrounded by a large area of farmland and forest, was used exclusively to house the campaign measurement team. This building stood with almost no other buildings within 300 m radius area. The instrument for determining hydroperoxides was located in an air conditioned room on the top floor, and the sampling inlet was mounted

www.atmos-chem-phys.net/8/6755/2008/ Atmos. Chem. Phys., 8, 6755–6773, 2008
on the rooftop. The space over the rooftop was very open to surroundings, thus made it possible to collect well mixed and representative air samples. Ambient air was drawn by a vacuum pump through a 6 m Teflon tube (1/4 inch O.D.) extending 1.5 m above the rooftop of the building, so that the air samples were taken about 12 m above the ground. The air flow rate was 2.7 slm (standard liters per minute), controlled by a mass flow controller. The air residence time in the inlet tubing was less than 2 s, and there was no filter in the inlet system. The air samples were collected in a thermostatically controlled glass coil collector, at a temperature of around 10°C. The stripping solution, acidified 18 MΩ water (H3PO4, pH 3.5) was delivered into the collector by an HPLC pump (Agilent 1050) at a rate of 0.2 mL min⁻¹ to collect hydroperoxides. The coil itself is about 30 cm long and the tube has an effective length of ~100 cm and 2 mm I.D. (Sauer et al., 1999). The scrubbing coil is similar to that used in earlier studies (Lazrus et al., 1986; Neeb et al., 1997; Sauer et al., 1999, 2001; Grossmann et al., 2003; François et al., 2005).

The collection efficiency of the coil was determined as follows. First, vapor containing H2O2 and MHP was generated by a saturated vapor generator (Lind and Kok, 1986; Li et al., 2004). The air stream flowed over the thermostatically controlled quartz fiber membrane (15 ±0.2°C), which was saturated by the standard solution, at a rate of 0.2 slm. Lind and Kok (1986) demonstrated that the air stream rate should be less than 1 slm in order to ensure Henry’s Law equilibrium. Second, additional pure air (2.5 slm) was added to the generated vapor of hydroperoxides via a three-port valve. Then the mixed air stream of standard gas of H2O2 and MHP was drawn into the scrubbing coil collector at a total flow rate of 2.7 slm under the conditions used for atmospheric measurement. Using a solution containing 2.4 × 10⁻³ M H2O2 and 3.5 × 10⁻⁶ M MHP, the levels of gaseous hydroperoxides in the standard gas were calculated to be ~1 ppbv for H2O2 and ~0.5 ppbv for MHP. The concentration of this standard gas was also determined using a Horibe tube in a cold trap of ethanol/liquid nitrogen at ~−90°C (Hewitt and Kok, 1991), for collection and for HPLC analysis (described below). The standard gas concentration determined by the cold trap method was consistent with the concentration calculated by Henry’s Law (H2O2=1.8×10⁵ M atm⁻¹ and H3MHP=5.7×10² M atm⁻¹ at 15°C, Sander et al., 2003). After collection, the stripping solution was analyzed by HPLC. The collection efficiency of the coil was estimated using the ratio of the measured concentration and the known concentration of the standard gas, with ≥98% for H2O2 and ~85% for MHP at 10°C. These values are in agreement with those of previous studies (Sauer et al., 1997, 2001; François et al., 2005). The heterogeneous decomposition of H2O2 and MHP in the coil was negligible under the experimental conditions, as proved by previous studies (Sauer et al., 1996, 2001).

After the sampled air passed through the coil collector, the stripping solution was removed from the separator using a peristaltic pump and immediately injected manually into the HPLC valve, from which 100 µL was analyzed by HPLC. Because of the lack of an auto-sampler for the HPLC analysis, the sample analysis was performed in a quasi-continuous mode with an interval of 20–60 min, and thus only a few samples were measured at night and in the early morning. Several rain samples were collected during a heavy shower using a glass funnel (diameter 10 cm) connected to a 5 m Teflon tube (1/8-inch O.D.), from the end of which the rainfall samples were collected and injected immediately into the HPLC column.

The HPLC was done with post-column derivatization using p-hydroxyphenylacetic acid (POPHA) and fluorescence detection. The basis of this method is to quantify the fluorescent dimer produced by the stoichiometric reaction of POPHA and hydroperoxides through catalysis (Gib et al., 1985; Hellpointner and Gib, 1989; Kurth et al., 1991; Lee et al., 1995; Sauer et al., 1996, 1997, 1999, 2001; Grossmann et al., 2003; François et al., 2005; Xu and Chen, 2005; Walker et al., 2006). The catalyst used in this study was Hemin (Xu and Chen, 2005; Chen et al., 2008). The mobile phase, controlled by the HPLC pump (Agilent 1200) at a constant rate of 0.5 mL min⁻¹, was a H3PO4 solution at pH 3.5 (Sigma-Aldrich, 85% for HPLC). The hydroperoxides were separated in a 5 µm reversed-phase C18 HPLC column (4.6 mm × 250 mm, ZORBAX, SB-Aq, Agilent), which was cooled to ~2°C to stabilize the hydroperoxides. After separation, the eluate was introduced into a 3 m Teflon coil at 42(±1)°C for post-column derivatization. The fluorescent reagent, 8 × 10⁻⁶ M Hemin (Fluka) and 8 × 10⁻⁵ M POPHA (ACROS ORGNICS), was adjusted to pH 10–11 with NH4Cl/NH4OH buffer solution. The flow rate of the fluorescent reagent was 0.2 mL min⁻¹. The fluorescence signal of the biphenyl derivative formed in the derivatization reaction was determined at wavelengths of λex=315 nm and λem=400 nm using a fluorescence detector (Agilent 1200).

Sample blanks were determined at least twice daily by measuring the stripping solution at the stripping solution outlet of the coil after stopping the air vacuum pump for 10 min. H2O2 was occasionally found in the blanks but only in trace amounts. When the pure synthetic air was introduced to the sampling inlet instead of the ambient air, our instrument did not detect any peroxides in the samples, indicating that the artefact production of peroxides within the sampling set-up was negligible. Multipoint calibration of the HPLC for analysis of hydroperoxides was performed weekly with H2O2, MHP and EHP standard solution in the range of 1×10⁻⁸ ~1×10⁻⁵ M, and single-point calibration was done three times a day with a mixing standard solution of H2O2, MHP and EHP. Organic hydroperoxides were identified by comparing the retention times with those of reference substances. The HPLC chromatogram shown in Fig. 1 illustrates the separation and retention times of a mixture of hydroperoxides during the measurements. The detection limit of H2O2, defined as three times the standard deviation of the
analytical blanks of Milli-Q water, was 0.012 \mu M \text{L}^{-1} using a 100 \mu L sampling loop. The detection limit of MHP, based on three times the standard deviation of the MHP standard solution (7.0 \times 10^{-7} \text{ M}), was 0.021 \mu M \text{L}^{-1}. These corresponded to the detection limits of about 20 pptv for \text{H}_2\text{O}_2 and 35 pptv for MHP in the gas phase under the sampling conditions mentioned above.

\text{H}_2\text{O}_2 was purchased from Sigma-Aldrich (35%), and fresh solutions were prepared by serial dilution of the 0.35% stock solution. Methyl hydroperoxide and ethyl hydroperoxide were synthesized from \text{H}_2\text{O}_2 and dimethyl sulfate or diethyl sulfate as described (Rieche and Hitz, 1929; Kok et al., 1995; Lee, et al., 1995). The purities of the synthesized MHP and EHP were proved to be higher than 97% using infrared spectrometry. The hydroxymethyl hydroperoxide (HMHP), 1-hydroxy-ethyl hydroperoxide (1-HEHP) were synthesized from aqueous \text{H}_2\text{O}_2 and formaldehyde or acetaldehyde (Rieche and Meister, 1935; Zhou and Lee, 1992; Lee et al., 1995). The concentrations of stock solutions and standard solutions were determined using KMnO$_4$ and KI/Na$_2$S$_2$O$_3$/starch every two weeks (Johnson and Siddigu, 1970; Mair and Hall, 1970). All reagents and standard solution were prepared with 18 M\Omega Milli-Q water (Millipore), and were stored at 4°C in a refrigerator.

### 2.2.2 Estimation of \text{SO}_2 interference

The \text{SO}_2 interference on the peroxides measurement should be clarified before interpreting the observations. We estimated this interference by two methods as following.

One estimation was made by a kinetics calculation. At 10°C temperature adopted in our sample collection, the Henry’s Law constants of \text{H}_2\text{O}_2 and \text{SO}_2 are 2.85 \times 10^3 \text{M atm}^{-1} and 2.32 \text{M atm}^{-1} respectively (Sander et al., 2003). The total amount of dissolved S(IV) always exceeds that predicted by Henry’s Law for \text{SO}_2 alone, so we calculate the effective Henry’s Law constant of \text{SO}_2 by the expression given in textbook (Seinfeld and Pandis, 1998), and obtain a value of 98 M atm$^{-1}$ at pH 3.5. Assuming the mixing ratios of \text{H}_2\text{O}_2 and \text{SO}_2 were 1 ppbv and 10 ppbv respectively in the sampled air, and then the equilibrium concentrations of \text{H}_2\text{O}_2 and \text{SO}_2 in collection solution at pH 3.5 are estimated to be 285 \mu M and 1 \mu M respectively in solution, on the basis of Henry’s Law. Then, the reaction rate of the dissolved \text{H}_2\text{O}_2 and S(IV) can be calculated to be $\sim 6.6 \mu \text{M s}^{-1}$ at pH 3.5. In the glass coil collector used in this study, the contact time of the sampled air with the collection solution was less than 10 s. After this time, the hydroperoxides-containing collection solution was separated from the sampled air, and was rapidly delivered into the HPLC system. This means that the \text{SO}_2 in sampled air can destroy \text{H}_2\text{O}_2 and influence \text{H}_2\text{O}_2 measurement only within this 10 s. On the basis of above calculation, we can give an estimation that, when 1 ppbv \text{H}_2\text{O}_2 and 10 ppbv \text{SO}_2 exists in the sampled air, about 66 \mu M of 285 \mu M \text{H}_2\text{O}_2 equilibrated in the collected solution will be destroyed by \text{SO}_2 during collection, indicating that the mixing ratio of \text{H}_2\text{O}_2 will be underestimated by 23%. More details regarding the interference of \text{SO}_2 in \text{H}_2\text{O}_2 measurement are as follows: when the \text{SO}_2 presents in sampled air with 2, 5, 10, 15, 20 and 30 ppbv respectively, the losses of \text{H}_2\text{O}_2 are estimated to be 5%, 12%, 23%, 35%, 46%, 70%, respectively.

Another estimation was made by a laboratory experiment. We studied the \text{SO}_2 interference using the Horibe tube, which was described above. The Horibe tube was added with 1 ml mixing standard solution containing $10^{-6}$ M \text{H}_2\text{O}_2, \text{MHP} and \text{EHP}, comparable to their concentrations in the atmosphere, following by cooling it at $\sim -90^\circ\text{C}$ and then introducing the \text{SO}_2-containing air flow. After the \text{SO}_2-containing air flow passed it for 30 min at the rate of 2.7 slm, the Horibe tube was taken out from the cold trap, and 2 ml of $10^{-3}$ M \text{H}_3\text{PO}_4 solution was added quickly to wash the inside of the Horibe tube, and then the washed sample solution was quickly analyzed by HPLC. In this procedure, the contact time of the hydroperoxides with the S(IV) in the solution was $\sim 5$ min at $\sim 10^\circ\text{C}$. We found that the losses of \text{H}_2\text{O}_2 are $\sim 10^\%$, $\sim 30^\%$ and $\sim 50^\%$ respectively when the \text{SO}_2 concentrations are 10 ppbv, 25 ppbv and 50 ppbv respectively. The losses of the organic hydroperoxides were lower than that of \text{H}_2\text{O}_2.

In summary, we suggest that the interference of \text{SO}_2 in the peroxides measurement will not be significant under a 15 ppbv level of \text{SO}_2, under our observation conditions.

### 2.3 Measurement methods for other trace gases

HO$_2$ radicals were measured at the Backgarden site by a laser-induced fluorescence instrument, operated by Forschungszentrum Juelich (FZJ). Briefly, ambient air is sampled continuously into a low-pressure detection chamber, where HO$_2$ is chemically converted to OH by reaction with added NO. The resulting OH is then detected by laser excited fluorescence at a wavelength of 308 nm. The instrument is calibrated by using the quantitative photolysis of water
vapour in synthetic air at 185 nm as a radical source. The accuracy of the measurements is estimated to be 20% for this campaign. Details of the instrument and its calibration can be found in Holland et al. (2003).

Semi-continuous measurements of WSOC were made by University of Tokyo (UT) using a particle-into-liquid sampler (PILS) followed by online quantification of TOC every 6 min using a total organic carbon (TOC) analyzer. Ambient aerosol was sampled at a flow rate of 16.7 L/min by the PILS, which used a steam saturator to grow the aerosol to sizes that can be collected by inertial impaction. The carbonaceous compounds in the liquid sample were then quantified online with the TOC analyzer. Details of the instrument can be found in Miyazaki et al. (2006).

The sulfate measurements were performed by the Aerodyne Aerosol Mass Spectrometer (AMS), operated by University of Tokyo (UT). The AMS can measure size-resolved chemical composition of ambient non-refractory (vaporized at 600°C under high vacuum) submicron aerosol for an integration time of 10 min. The AMS consists of a particle sampling inlet with cutoff size 1.0 μm, a particle time-of-flight (PTOF) chamber, and a vaporizer/ionizer that is interfaced to a quadrupole mass spectrometer (QMS). Details of the instrument can be found in Takegawa et al. (2005).

The continuous measurements of 47 C2–C12 VOC species, including alkanes, alkenes and aromatic hydrocarbons, were performed hourly using an automated online Gas Chromatography-Flame Ionization Detector (GC-FID) system from the Research Center for Environmental Changes (RCEC), Academia Sinica of Taiwan. The detection limit for the measured VOC species ranged from 0.002 to 0.098 ppbv. The calibrations of the instrument were carried out before and after the campaign using a standard gas prepared by the gravimetric method (Spectra gases, Branchburg, NJ, USA), which contained 50 target species with mixing ratios of 3–15 ppbv. It was found that the precisions for the measured VOC species were usually better than 2%. Details of the GC-FID system and analytical procedures employed in this study can be found in Wang et al. (2004).

The SO2 was determined by Peking University (PKU) using SO2 Analyzer (Thermo, Model 43C) with a time resolution of 1 min. The data of CO, O3 and NO2 used in this study were obtained from the combined data set of PKU, UT and FZJ. During the period we discuss in this study, the CO was measured by a CO Analyzer (Thermo, Model 48C) with a time resolution of 1 min, operated by PKU, and a non-dispersive infrared absorption (NDIR) instrument with an integration time of 1 min (Model 48, TECO), operated by UT (details of the instrument described by Takegawa et al., 2006), and the O3 was mainly measured by a O3 Analyzer (Thermo, Model 49C) with a time resolution of 1 min operated by PKU. NOx and NO2 were measured using a NO-NOx chemiluminescence detector combined with a photolytic converter and a gold tube catalytic converter (Takegawa et al., 2006). NOx compounds were catalytically converted to NO on the surface of a gold tube heated at 300°C. The photolytic converter system used for the NO2 measurement was manufactured by the Droplet Measurement Technologies, Inc., USA.

3 Results and discussion

3.1 General observations

A total of 354 air samples were characterized using the scrubbing coil collector from the 19th to the 30th of July 2006 during the PRIDE-PRD’06 campaign. The major hydroperoxide present in the air samples collected at the Backgarden site was H2O2 with mixing ratios between below the detection limit (20 pptv) and 4.6 ppbv, and MHP with mixing ratios between <0.35 pptv (d.l.) and 0.8 ppbv. The organic peroxides BHMP and PAA were often detected, and GMHP, 1-HEHP and EHP were occasionally detected, but all these species were present at only several-decade pptv level under these experimental conditions. In order to calculate the mean of the observed mixing ratios, any value below the detection limit was treated as zero. With regard to all samples, the mean (and standard deviation) mixing ratios during the daytime (08:00–20:00 LT) were 1.26±0.42 ppbv for H2O2 and 0.28±0.10 ppbv for MHP. The mean values at night (20:00–02:00 LT) were 0.74±0.62 ppbv for H2O2 and 0.19±0.10 ppbv for MHP. The mixing ratios of H2O2 and MHP are in agreement with those reported in the literature with H2O2 ranging between 0.5 ppbv and 5 ppbv and MHP ranging between several pptv and 2.7 ppbv worldwide respectively (Hellpointner and Gäb, 1989; Hewitt and Kok, 1991; Das and Aneja, 1994; Watkins et al., 1995a, b; Jackson and Hewitt, 1996; Sauer et al., 1997, 2001; O’Sullivan et al. 1999; Morgan and Jackson, 2002; Moortgat et al., 2002; Grossmann et al., 2003; Lee et al., 1993, 1995, 1998, 2000, 2008; François et al., 2005; Xu and Chen, 2005; Walker et al., 2006; Kim et al., 2007).

Temporal profiles of the H2O2 and MHP mixing ratios for the time of the campaign are shown in Fig. 2. The maximum mixing ratio of H2O2 and MHP was found on 19 July, and this will be discussed in detail later. On sunny days with low levels of NOx and SO2, H2O2 showed pronounced diurnal variations, with peak mixing ratios in the afternoon (12:00–18:00 LT) and low values at night and in the early morning. Sometimes, a second peak occurred in the evening between 20:00 and 02:00 LT. The diurnal variation of MHP is consistent with, but less pronounced than, that of H2O2. The general diurnal cycle of H2O2 observed at Backgarden was similar to that observed in earlier studies (Sauer et al., 2001; Grossmann et al., 2003). Over the 13 days of measurement, HMHP was detected in only a few samples; probably resulting from the heterogeneous decomposition of HMHP at glass surfaces during sampling (Neeb et al., 1997; Sauer et al. 2001).
Fig. 2. Temporal profiles of atmospheric H$_2$O$_2$ and MHP mixing ratios from 19 to 30 July 2006 at Backgarden.

Fig. 3. Hourly averaged diurnal cycle for H$_2$O$_2$ (black circle) and MHP (gray triangle) at Backgarden from 19 to 30 July 2006 where vertical bars show the standard deviation.

Figure 3 depicts the hourly averaged mixing ratio profiles of H$_2$O$_2$ and MHP with the vertical bars showing the standard deviation of the measured values. A similar averaged diurnal profile for H$_2$O$_2$ has been reported (Das and Aneja, 1994; Sauer et al., 2001). The concentration of H$_2$O$_2$ began to rise in the morning (∼10:00 LT) and reached a maximum mixing ratio at 13:00 LT. The factors responsible for the H$_2$O$_2$ diurnal variation are discussed in detail in Sect. 3.2.2. The level of H$_2$O$_2$ remains relatively high in the afternoon and the mixing ratio decreased slowly from sunset to 24:00 LT. The diurnal profile of MHP is largely coincident with that of H$_2$O$_2$ in the daytime, but remained at an almost identical level after sunset. This slower loss of MHP at night can be explained by its lower level of solubility (H$_{H_2O_2}$/H$_{MHP}$=∼260, at 298 K, Sander et al., 2003).

The diurnal profiles of the meteorological parameters and other trace gases measured at Backgarden from 19 to 30 July are shown in Fig. 4. As a comparison, H$_2$O$_2$ and MHP are also shown in this figure. It is seen that when the SO$_2$ level remained high, >15 ppbv, in the morning or elevated during the day, H$_2$O$_2$ were detected in very low concentrations. These low concentrations resulted partly from the potentially significant interference of SO$_2$ in the H$_2$O$_2$ measurement. Meanwhile, considering the fact that the loss of H$_2$O$_2$ due to SO$_2$ reaction in the atmosphere would also be amplified by elevated SO$_2$, we suggest that the diurnal cycle of H$_2$O$_2$ shown in Fig. 2 and Fig. 4 could reflect the variation trend of H$_2$O$_2$ in the atmosphere to some extent, although the uncertainty of H$_2$O$_2$ quantification occurs. However, it is noteworthy that at the SO$_2$ level of <10 ppbv, the condition for most of the daytime during the measurement period, the measurement of H$_2$O$_2$ were not significantly influenced, that is, most of our data were not significantly influenced by interference of SO$_2$ and are reliable.

With regard to the meteorological conditions and levels of hydroperoxides, three distinct periods could be distinguished. (i) At the beginning of the measurement, 19–21 July, days were sunny with slight breeze, and hydroperoxides exhibited high mixing ratios during the day. (ii) The second period, 23–26 July, was influenced by typhoon Kaemi, which came across most of the PRD and resulted in heavily polluted conditions in the central and eastern PRD (Z. B. Yuan, 2007, personal communication). During this period, high levels of hydroperoxides were also observed at Backgarden on 24 and 25 July, two sunny days. (iii) During the last days of the campaign, 27–30 July, the local weather conditions were cloudy and rainy, and daytime values of hydroperoxides were low.

The low daytime average H$_2$O$_2$ values probably result from several factors, and the most important one is that the weak photochemical activity on cloudy days produces fewer
HO$_x$ radicals compared to sunny days, resulting in low-level production of hydroperoxides. Moreover, the high levels of NO$_x$ will significantly suppress the formation of hydroperoxides by consuming their precursors, peroxo radicals. Additionally, efficient scavenging of H$_2$O$_2$ on wet surfaces (leaves and fog droplets) and water-covered aerosols, in particular with a high level of SO$_2$ and high relative humidity conditions, should partly account for the low levels of H$_2$O$_2$.

3.2 Photochemistry on sunny days

3.2.1 Pattern of hydroperoxides and their precursors

The meteorological conditions during 19–21 July at Backgarden can be treated as identical. On these three sunny days, the maximum temperature was 35°C, and the relative humidity decreased from ~90% in the early morning to ~60% at noon. After reaching a minimum level of ~45% in the afternoon, relative humidity increased gradually until the next morning. The wind speed was steady at around 0–3 m/s. The wind direction on 19 July turned clockwise via southeast in the morning to southwest at noon and back to southeast...
gradually in the late afternoon, and then remained southeast during the night. A similar pattern of wind direction was observed on the next two days.

The maximum mixing ratio of H$_2$O$_2$ and MHP was measured on 19 July, a sunny day with a slight breeze. As Fig. 4 shows, the concentration of NO$_x$ stayed very high from night to the early morning, with 32 ppbv at 08:15 LT. The high NO$_x$ mixing ratio might have been caused by the accumulation of surface emission below the nocturnal inversion layer. After 08:15 LT, the concentration of NO$_x$ decreased rapidly to ~3 ppbv at 10:30 LT, and remained at a relatively low level (1~3 ppbv) until sunset. The rapid drop of NO$_x$ probably resulted from the growing height of the mixing layer and changes in the wind direction. On the morning of 19 July, the wind direction turned clockwise via northeast (06:00 LT) to east (08:00 LT), and then to south (09:30 LT). Moreover, the chemistry conversion of NO$_x$ to HNO$_3$ might have been caused by the accumulation of surface emission below the nocturnal inversion layer.

Xing et al. (2005) suggested that a value of a NO$_x$/benzene of <0.5 is indicative of photochemically aged air due to the shorter atmospheric lifetime of toluene compared to benzene. Therefore, the high levels of hydroperoxides in this period were thought to be due to a combination of photochemically aged air with very high levels of HO$_x$, which were also measured at Backgarden. The HO$_x$ concentration can be used as an indicator for photochemical effective radiation. At Backgarden, the average NO mixing ratio at 10:30–14:00 LT and 14:00–18:00 LT on 19–21 July were ~280 ppbv and ~80 ppbv respectively, and the mixing ratio of HO$_2$ radicals was ~2×10$^9$ molecule/cm$^3$ at noon. Using the HO$_2$ average concentration, it can be obtained from a simple calculation that when the concentration of NO is ~65 ppbv, the reaction rate of NO–HO$_2$ and HO$_2$–HO$_2$ could be about equal. Hence, high levels of hydroperoxides detected in the daytime may be attributed to the moderate level of NO$_x$ and exceptionally high mixing ratio of HO$_2$ radicals produced by oxidation of VOC and CO at Backgarden.

### 3.2.2 Kinetics analysis

In general, j(NO$_2$) can be used as an indicator for photochemically effective radiation. At Backgarden, j(NO$_2$) usually began to rise after 06:00 (LT), reached maximum values of ~8×10$^{-3}$ s$^{-1}$ at noon and then returned to near-zero after 19:00 (LT) (personal communication by Bohn, B., Forschungszentrum Juelich). During 19–21 July, the maximum mixing ratios of H$_2$O$_2$ were observed during the daytime, and the diurnal variation of H$_2$O$_2$ was generally similar to that of j(NO$_2$), but the peak values were ~2–3 h later. Generally, the photo-oxidant formation began about 3 h after the increase of radiation. The peak time of H$_2$O$_2$ approached that of O$_3$ on 19 and 20 July, and the diurnal profiles of these two species were similar. Additionally, peroxy acetic acid (PAA), which is produced mainly by photo-oxidation of acetone and PAN, was often detected on 19–21 July. On the basis of this evidence, we can infer that H$_2$O$_2$ and MHP were produced, to a large extent, in the daytime by the local photochemical process during the three days.

Even more direct evidence of the photochemical formation of hydroperoxide can be obtained from the diurnal profiles of HO$_2$, which were also measured at Backgarden. The HO$_2$ concentration can be used to calculate the chemical production rate of H$_2$O$_2$. The mixing ratios of HO$_2$ and H$_2$O$_2$ measured on 21 July are shown in Fig. 5. The mixing ratios of H$_2$O$_2$ and HO$_2$ are almost zero at the high concentration of NO$_x$ before 09:30 (LT). The sharp increase of H$_2$O$_2$ at about 09:45 (LT) on 21 July coincides

As described previously, the formation of hydroperoxide can be represented by Reactions (R1) and (R2) ($k_1$ is taken from Sander et al. (2003), at 298 K):

$$
\text{HO}_2 + \text{HO}_2 \xrightarrow{k_1} \text{H}_2\text{O}_2 + \text{O}_2
$$

$$
\frac{\text{HO}_2 + \text{HO}_2 + (M) \xrightarrow{k} \text{H}_2\text{O}_2 + (M)}{M} \quad \text{(R2)}
$$

($M$ is air; the calculation of $k$ should take into account the pressure dependence and the temperature dependence and water vapor dependence.)

However, the NO reaction with peroxy radicals will compete with the formation of hydroperoxides, since the reactions of NO with peroxy radicals are faster than recombination reactions of peroxy radicals (Lee et al., 2000). At Backgarden, the average NO mixing ratio at 10:30–14:00 LT and 14:00–18:00 LT on 19–21 July were ~280 ppbv and ~80 ppbv respectively, and the mixing ratio of HO$_2$ radicals was ~2×10$^9$ molecule/cm$^3$ at noon. Using the HO$_2$ average concentration, it can be obtained from a simple calculation that when the concentration of NO is ~65 ppbv, the reaction rate of NO–HO$_2$ and HO$_2$–HO$_2$ could be about equal. Hence, high levels of hydroperoxides detected in the daytime may be attributed to the moderate level of NO$_x$ and exceptionally high mixing ratio of HO$_2$ radicals produced by oxidation of VOC and CO at Backgarden.
with the decrease of the NOx, SO2 and CO mixing ratios, which might be explained by vertical exchange and changes in the wind direction. In the morning, the wind direction turned clockwise via northeast (07:30 LT) to southeast (09:30 LT). During 10:00–12:00 LT, H2O2 continued to rise at a rate of \( \sim 0.81 \text{ ppbv h}^{-1} \), and the chemical production rate of H2O2 was \( \sim 0.74 \text{ ppbv h}^{-1} \), as determined from the HO2 concentration of \( \sim 8.9 \times 10^6 \text{ molecule cm}^{-3} \). The calculation adopts the expressions recommended by Stockwell (1995), and the HO2 concentration and temperature uses the average value during the period, resulting in a rate coefficient of \( 6.5 \times 10^{-12} \text{ cm}^3 \text{ molecule}^{-1} \text{s}^{-1} \) at 60% relative humidity. This indicates that most of the H2O2 increase was produced by in situ formation and the rest might be attributed to the net effect of vertical mixing. As shown in Fig. 5, the diurnal cycles of HO2 and H2O2 were, in general, consistent in the afternoon until 17:00 (LT); after that, the level of H2O2 had a weak correlation with that of HO2. It is worth noting that the relative humidity rose rapidly after 17:00 LT, while the concentration of NO remained low (\( \sim 30 \text{ pptv} \)) in the evening of 21 July.

The decrease of H2O2 during the late afternoon until the night might be attributed to the following reasons. First, considering the high solubility of H2O2 (Lind and Kok, 1994; O’Sullivan et al., 1996), the observed low levels can be explained by increased relative humidity (\( \sim 80\% \) at 21:00 LT), which results in greater wet deposition of H2O2 at night than during the daytime. Secondly, the dry deposition of hydroperoxides on the Earth’s surface will become very pronounced under a shallow inversion and at a low wind speed. The wind speed in the evening on 21 July was \( \sim 1 \text{ m/s} \); therefore, dry deposition on the surface might have acted as an important sink for loss of H2O2. Moreover, Walcek (1987) and Wesley (1989) have found that the deposition rate of H2O2 over trees is much higher than in the free troposphere. Hence, the low mixing ratios of H2O2 might be due, in part, to the deposition on the leaves of the dense forests surrounding the observation site. Furthermore, when the temperature decreased during the night, Henry’s Law constant of H2O2 will increase, resulting in a removal of H2O2 from the gas phase into the liquid phase. As a result, the vast lake adjacent to the observation site might be substantially responsible for the decrease of H2O2. In addition, resulting from the limitation of the collection method, the loss of H2O2 by elevated SO2 during collection could also be an important reason for the decrease of gaseous H2O2 in the late afternoon.

### 3.2.3 Impact of local meteorology on hydroperoxides

The two sunny periods discussed here suggest that the hydroperoxide formation at Backgarden is, to a large extent, a local phenomenon. High levels of hydroperoxides were observed in the two sunny periods between 19–21 July and 24–25. The mixing ratios of hydroperoxides were similar in the two periods. Moreover, the diurnal variation of H2O2 showed a positive correlation with O3 on 24 July, as shown in Fig. 4, with the peak time of H2O2 2–3 h later than that of O3. A ratio of toluene/benzene of <0.5 was observed between 12:00 LT on 24 July to 21:00 LT on 25 July, with a few exceptions in the early morning of 25 July. This indicates that during that time the air at Backgarden influenced by the typhoon front was photochemically aged. All the evidence indicates that local photochemical activity contributed substantially to the levels of hydroperoxides during 24 and 25 July.

It is worth noting that the dominant wind directions in the two sunny periods were opposite. As mentioned previously, southeasterly winds prevailed at the observation site during 19–21 July. On 24 and 25 July, the wind direction at Backgarden was northerly and veered to northwesterly in the afternoon, consistent with that of back trajectories obtained from NOAA (www.arl.noaa.gov). The wind speeds measured during the daytime of these two periods were similar, at \( \sim 2 \text{ m/s} \), ensuring transport of air masses over distances \( \sim 30 \text{ km} \) between sunrise and the maximum observed photo-oxidant values. This suggests that the levels of hydroperoxides at Backgarden were not influenced by transport at low wind speed. Thus, much of the variation of hydroperoxide mixing ratios observed at Backgarden on these sunny days can be attributed, to a large extent, to the local photochemical drive.

### 3.3 Rain

The heavy shower that started at 21:20LT on 25 July and lasted for 40 min was brought by the typhoon Kaemi. At 17:00 LT, the wind direction turned from north to northwest, and the mixing ratios of NOx, SO2 and CO began to rise, reaching 19 ppbv, 9 ppbv and 1.6 ppmv, respectively, at 21:00 LT, while O3 decreased from 54 ppbv to 8 ppbv, as shown in Fig. 4. At the same time, the levels of hydroperoxides decreased rapidly, i.e. \( \text{H}_2\text{O}_2 \) went from 3.2 ppbv to 0.9 ppbv and MHP went from 0.6 ppbv to 0.3 ppbv. These
changes were interrupted at 21:20 LT when there was a heavy shower at Backgarden. When the rain began to fall, the temperature at ground level was 302 K. The shower lasted for ~40 min with lightning activity. During the shower, three rainfall samples were collected and analyzed immediately. The maximum concentration of H$_2$O$_2$ in the rain samples, 21 µmol/L, was detected at the beginning of the shower. This concentration is within the range from 0.1 µmol/L to 300 µmol/L reported by earlier studies (Hellpointner and Gäb, 1989; Jacob et al., 1990; Hewitt and Kok, 1991; Sauer et al., 1997; Morgan and Jackson, 2002). Moreover, MHP, which was seldom observed in rain samples (Hellpointner and Gäb, 1989; Pena et al., 2001; Reeves and Penkett, 2003), was detected in the rain samples at Backgarden at a concentration of 1.1 µmol/L. This value may represent the concentration of MHP in cloud water. If we assume that the MHP value in the gas phase at the height of the cloud base was the same as that detected at ground level, ~0.5 ppbv, the equilibrium concentration of MHP in cloud water is estimated to be 0.2 µM, on the basis of Henry’s Law (H$_{MHP}$=4.16×10$^2$ M atm$^{-1}$, 293 K, Sander et al., 2003). This estimated value is much smaller than the concentration of MHP detected in the rainwater, which implies a higher gas-phase level of MHP in the clouds compared to that at ground level. Similarly, this higher concentration above PBL (1~2 km) can be estimated by Henry’s Law. The ambient temperature will decrease 6~7 K when the altitude increases by 1 km; thus, the temperature at the height of the cloud base can be estimated to be ~293 K, while the temperature at ground level was 302 K. According to the concentration of MHP detected in the rain (1.1 µmol/L), the gas-phase MHP mixing ratio above PBL was ~2.6 ppbv. This estimated value is slightly higher than those reported for earlier field studies in which MHP was detected directly by aircraft (O’Sullivan et al., 1999; Lee et al., 2000). MHP may be of great importance in the redistribution of OH radicals along with the driving force of atmospheric chemistry (Wennberg et al., 1998; Cohan et al., 1999; Ravetta et al., 2001; Mari et al., 2000). Our measurement may be new evidence for the existence of high mixing ratios of MHP above the PBL.

The levels of hydroperoxides after the shower lend support to the deduction that high mixing ratios of hydroperoxides occur in the PBL. H$_2$O$_2$ and MHP exhibited relatively high mixing ratios of 2.1 ppbv and 0.64 ppbv, respectively, immediately after the shower; meanwhile, the mixing ratios of NO$_x$, SO$_2$ and CO decreased to relatively low values due to the dilution and scavenging effects, as shown in Fig. 4. The mixing ratios of hydroperoxides after the shower were even higher than they were before the shower. Considering the much higher solubility of H$_2$O$_2$ than that of NO$_x$, SO$_2$ and CO, we suggest that vertical convection might contribute significantly to the increased H$_2$O$_2$ and MHP mixing ratios, for the following two reasons. First, the air mass above and in the boundary layer may be carried down to the land surface when rain falls. As a result, the gas-phase H$_2$O$_2$ above the PBL that was not washed out by the shower might affect the mixing ratio at low altitudes. Second, the falling rain and rainwater on the ground (e.g. on the leaves of plants) might release H$_2$O$_2$ and MHP into the gas phase during and after the shower, because of the decrease of Henry’s Law constants due to the increase of temperature with descending altitude. In addition, owing to the low level of solubility and its estimated 2–3 days atmospheric lifetime (Cohan et al., 1999; Wang and Chen, 2006), a fraction of the increased MHP might be introduced partly by the advection of typhoon from other regions. Moreover, although the measurement of VOC was interrupted in the hours following the shower, the low mixing ratio of alkenes (~2 ppbv) at around 21:20 LT indicated that the ozonolysis of alkenes might have a minor impact on the level of hydroperoxides during the shower.

Overall, this measurement of hydroperoxides during the shower may provide evidence for the high mixing ratio of hydroperoxides, particularly MHP, above the boundary layer. This mixing ratio of MHP might potentially influence the redistribution of HO$_x$ and RO$_x$ radical in the PRD on a regional scale.

### 3.4 Formation of hydroperoxides by ozonolysis

The ozonolysis of alkenes (e.g. isoprene, terpenes, ethene, propene and isobutene) can produce a variety of peroxides (Gäb et al., 1985, 1995; Becker et al., 1990, 1993).

It is proposed that ozonolysis proceeds by the initial insertion of the ozone into the double bond forming a primary ozonide, and decomposes to form excited Criegee intermediates (ECI) [R$_1$R$_2$COO]* and a carbonyl compound (Gäb et al., 1985). ECI are biradicals with excess energy, and some of them will become stabilized Criegee intermediates (SCI) R$_1$R$_2$COO by interaction with the medium, and the SCI can react further to produce hydroperoxides. Recent laboratory studies have revealed that R$_1$R$_2$C(OH)OOH can be formed by the reaction of SCI with water vapor (Horie et al., 1994; Neeb et al., 1997; Sauer et al., 1999; Valverde-Canossa et al., 2001). This R$_1$R$_2$C(OH)OOH decomposes primarily to H$_2$O$_2$ and a carbonyl compound R$_1$COR$_2$, as shown in the following reactions (R$_1$ and R$_2$ are alkyl groups):

\[
R_1R_2COO + H_2O \rightarrow R_1R_2C(OH)OOH \quad (R3)
\]

\[
H_2COO + H_2O \rightarrow HOCH_2OOH(R_1=H, R_2=H) \quad (R4)
\]

\[
R_1R_2C(OH)OOH \rightarrow R_1COR_2 + H_2O_2 \quad (R5)
\]

There is some evidence that H$_2$O$_2$ and MHP were formed in the evening. As shown in Fig. 6, a high H$_2$O$_2$ mixing ratio was detected after sunset (19:20 LT) on 24 July; in particular, a second peak (~1.9 ppbv) was observed during the evening. Relatively high mixing ratios of alkenes (~8 ppbv), particularly isoprene (~5 ppbv), were detected during the evening on 24 July, compared to the other nights. The mixing ratio of...
H$_2$O$_2$ at 21:00 LT was about half of the maximum value observed during the daytime. However, the level of HO$_2$ at this time was $\sim 3 \times 10^5$ molecules cm$^{-3}$, only $\sim 13\%$ of the maximum value observed at midday, as shown in Fig. 6. This high level of H$_2$O$_2$ production cannot be attributed to only the recombination of HO$_2$ radicals, suggesting that the formation via the ozonolysis of alkenes under moist atmospheric conditions (70\% relative humidity) contributes substantially to the production H$_2$O$_2$ during the evening. Further evidence for this pathway comes from the fact that HOCH$_2$OOCH$_2$OH (BHMP) was observed for a considerable length of time only during the night of 24 July. The apparent precursor of BHMP is HMHP (Gäb et al., 1985), which is a unique product of the ozonolysis of exocyclic biogenic alkenes (Valverde-Canossa et al., 2001). HMHP is formed by CH$_2$OO biradicals, which are produced in the ozonolysis of terminal alkenes, as shown in Reaction (R4), while the formation of BHMP can be expressed by Reaction (R6):

$$\text{HOCH}_2\text{OOH} + \text{HOCH}_2\text{OOH} \rightarrow \text{HOCH}_2\text{OOCH}_2\text{OH} + \text{H}_2\text{O} \quad \text{(R6)}$$

Therefore, the reaction of alkenes with O$_3$ can be suggested as a source of hydroperoxides at night at the Backgarden site. Grossmann et al. (2003) proposed that the ozonolysis of alkenes was a source of H$_2$O$_2$ at night at Pabstthum, Germany.

It is worth noting that MHP had a diurnal profile similar to that of H$_2$O$_2$ in the evening at Backgarden, and MHP also exhibited a second peak 0.4 ppbv at night on 24 July. This level of MHP at night was much higher than those reported for other continents (Hellpointer and Gäb, 1989; Jackson and Hewitt, 1996; Sauer et al., 2001; Grossmann et al., 2003; Walker et al., 2006).

From Fig. 4, it can be seen that relatively high levels of hydroperoxides were detected in the evening of 20 and 21 July, which coincided with the increase of C$_3$-C$_5$ alkenes, especially isoprene, measured in the same time periods. Moreover, BHMP was also observed in the morning and the afternoon of 25 July in addition to the afternoon and night of 24 July. Noticeably, when BHMP was detected, the mixing ratios of C$_3$-C$_5$ alkenes, particularly isoprene, propylene, 1-butene and 1-pentene were high, as shown in Fig. 4. Thus, we suggest that high levels of alkenes, especially isoprene, are favorable for the formation organic hydroperoxides. Besides BHMP, the relatively high level of PAA was often detected on 23–25 July in addition to 19–21 July mentioned in Sect. 3.2.2. The maximum mixing ratio of PAA during the observation was observed in the afternoon on 24 July, on which PAN also exhibited a high level of $\sim 3$ ppbv (B. Wang, 2008, personal communication, Peking University). Hence, we suggest that the high levels of PAN and strong solar radiation in the daytime could favor the formation of PAA. It is noteworthy that PAA was often detected after sunset on 20–21 and 23–25 July, implying that PAA might also be formed in the absence of photo-oxidation. However, to our best knowledge, PAA has not been reported as a product of the ozonolysis of alkenes, and the relevant mechanism is unclear and need further investigation. The solubility of BHMP is higher than H$_2$O$_2$ (H$_2$BHMP$=6.0 \times 10^5$ M atm$^{-1}$, 293 K), thus, it could enter into the aqueous phase easily. BHMP could easily decompose and convert to H$_2$O$_2$, thus it might take part in the formation of sulfate and SOA in the form of H$_2$O$_2$. PAA is an important acid in its own right and could decompose easily into acetic acid and H$_2$O$_2$ under typical tropospheric conditions. As a result, PAA could play a role in acid deposition both as an oxidant and an acid.

3.5 Hydroperoxide contribution to aerosols

3.5.1 Role of hydroperoxides in the formation of secondary sulfate

Atmospheric aerosols are responsible for the deterioration of air quality in industrialized areas and adversely affect human health and welfare. A major component of aerosols in North America, Europe, and Asia is secondary sulfate resulting from the atmospheric oxidation of anthropogenically emitted sulfur dioxide (SO$_2$) (US Environmental Protection Agency, 2001). Therefore, the oxidants and oxidation processes involved in the formation and growth of secondary sulfate are important subject in need of further study, especially when taking into account the long-range transport of anthropogenic sulfate aerosols (Perry et al., 1999). During the PRIDE-PRD’06 campaign, sulfate present in the aerosol phase was determined to be a major component, 10–60\%, of PM$_{2.5}$ mass (S. Guo, 2008, personal communication, Peking University). The main oxidation process for SO$_2$ in the atmospheric gas phase is its reaction with OH radicals.

However, it is suggested that the aqueous phase reaction with H$_2$O$_2$ and O$_3$ is the main route for SO$_2$ oxidation. The aqueous-phase oxidation with H$_2$O$_2$ accounts for 60–80\% of the total oxidation of SO$_2$ in the atmosphere, especially when the pH is $< 4.5$ (Penkett et al., 1979; Calvert et al.,
follows (Hoffmann and Edwards, 1975):

\[
\text{SO}_2 + \text{H}_2\text{O} \leftrightarrow \text{HSO}_3^- + \text{H}^+ \quad (\text{R7})
\]

\[
\text{HSO}_3^- + \text{H}_2\text{O} \rightarrow \text{HSO}_4^- + \text{H}_2\text{O} \quad (\text{R8})
\]

\[
R_a = -d[S(IV)]/dt = k_8[H^+][S(IV)] \quad (\text{M s}^{-1}) \quad (1)
\]

\[
R_a' = 10^{-6}L R_a \quad (\text{mol (L of air)}^{-1} \text{s}^{-1})
\]

\[k_8=7.5\pm1.6\times10^7 \text{M}^{-1} \text{s}^{-1} \text{ at 298 K.} \quad (\text{Seinfeld and Pandis, 1984})
\]

While the total amount of dissolved S(IV) always exceeds that predicted by Henry’s Law for SO\(_2\) by hydroperoxides in the aqueous phase are as follows (Hoffmann and Edwards, 1975):

\[
\text{Sulfuric acid by hydroperoxides in the aqueous phase are as}
\]

\[
\text{increased at a rate of } 10^{10} \text{ppbv for SO}_2, 2.8 \text{ppbv for H}_2\text{O}_2 \text{ and } 50 \text{ppbv for O}_3, \text{ the sulfate production rate is } 1.4\times10^{-12} \text{mol m}^{-3} \text{s}^{-1} \text{ in the aqueous phase (using Eq. (1) and (2)) and } 3.5\times10^{-12} \text{mol m}^{-3} \text{s}^{-1} \text{ in the gas phase, resulting in a combined sulfate production in both phases of } \sim4.9\times10^{-12} \text{mol m}^{-3} \text{s}^{-1}. \text{ This estimated sulfate production is much smaller than the measured value of } 1.7\times10^{-11} \text{mol m}^{-3} \text{s}^{-1}, \text{ which indicates that other processes are responsible for the missed source of sulfate. It is worth noting that the calculated production of sulfate mentioned above includes only the production in the gas phase and in the aqueous bulk of droplets. The heterogeneous chemistry on the surface of droplets and aerosols is potentially important (Li et al., 2006, 2007; Ammann and Pöschl, 2007; Pöschl et al., 2007; Chen et al., 2008), but it is not taken into account in the above estimation. Jayne et al. (1990) observed that the uptake of SO\(_2\) into water droplets was faster than predicted on the basis of the known kinetics in bulk solution, and they suggested that a surface complex was formed between SO\(_2\) and H\(_2\)O at the interface. Vácha et al. (2004) suggested that the concentration of H\(_2\)O\(_2\) is increased in the interfacial region by \sim50\% compared to the bulk. Chung et al. (2005) pointed out that salts containing ammonium ions were found to increase the solubility of H\(_2\)O\(_2\) by up to a factor of two compared to pure water. Hasson and Paulson (2003) found that the concentration of H\(_2\)O\(_2\) within aerosols was of the order of \sim10^{-3} \text{M}, which is one order of magnitude higher than the expected concentration based on the solubility of H\(_2\)O\(_2\) in liquid water (\sim1\times10^{-4} \text{M}). Moreover, Chen et al. (2008) recommended that the interfacial reaction should be taken into account in the generalized aqueous phase especially for a rapid reaction. Combining all these intriguing hints with our estimation, we suggest that the surface heterogeneous phase reaction, here,
the heterogeneous reaction of SO2 with H2O2 might make a substantial contribution to sulfate production. Clearly, the mechanism, kinetics parameters and yield of sulfate formation regarding the heterogeneous reactions need further investigation.

Recent studies have revealed that the enhanced acidity of the aerosol can catalyze particle-phase heterogeneous reactions of atmospheric organic carbonyl species (Jang et al., 2002, 2003; Iinuma et al., 2004). The reactions of SO2 with hydroperoxides produce sulfate, and provide hydrogen ions continuously for heterogeneous reaction systems.

3.5.2 Contribution to the formation of secondary organic aerosol (SOA) in aerosols

SOA formed through oxidation of atmospheric VOC contributes to the global aerosol burden through both biogenic and anthropogenic precursors. The biogenic fraction of SOA contributes the most, with estimates varying between 8 Tg yr⁻¹ and 40 Tg yr⁻¹ (Penner et al., 2001). Recent laboratory studies have revealed that acid-catalyzed heterogeneous reaction of isoprene and its gas-phase oxidation product with hydrogen peroxide lead to the formation of SOA (Claeys et al., 2004b; Böge et al., 2006; Kroll et al., 2006). This new route may explain the formation of water-soluble organic compounds (WSOC), which include hydroxyl and/or carboxyl functional groups and represent a considerable fraction of the SOA (Saxena and Hildemann, 1996).

Some evidence from the PRIDE-PRD’06 study implies that a negative correlation might exist between the observed hydroperoxides and the concentration of WSOC. Fig. 8 shows the measured concentrations of H2O2 and WSOC, and it can be seen that the diurnal variations of the two kinds of species are generally opposite.

It is generally accepted that the formation of SOA from biogenic hydrocarbons emitted by terrestrial vegetation is via gas-phase photochemical reactions followed by gas-to-particle partitioning (Seinfeld and Pandis, 1998). In the atmosphere, hydroperoxides and WSOC are competitive in their formation reactions, involving the intermediates R1R2COO and HO2 radicals. Additionally, WSOC may be produced by acid-catalyzed heterogeneous oxidation with hydrogen peroxide as reported (Claeys et al., 2004b; Böge et al., 2006; Kroll et al., 2006), resulting in the consumption of hydrogen peroxide. Thus, a negative correlation of atmospheric hydrogen peroxide with aerosol-phase WSOC could be expected to some extent. Recently, the laboratory study revealed that the aqueous-phase ozonolysis of isoprene and its gas-phase oxidation product may serve as a potentially important route for the formation of oxidants, including H2O2 (Chen et al., 2008). The field evidence indicated that the sampled particles are capable of generating H2O2 in aqueous solution (Arellanes et al., 2006). Although it is difficult to distinguish quantitatively the contribution of gas-phase H2O2 and H2O2 generated in aqueous phase, our observations might provide some evidence that atmospheric H2O2 potentially contributes to the formation of WSOC and a negative correlation might exist between the two kinds of species, as shown in Fig. 8. Combining this H2O2-WSOC negative correlation and the H2O2-sulfate negative correlation as shown in Fig. 7, the results of our field observations seem to imply that hydrogen peroxide and sulfuric acid potentially play a potential important role in the formation of WSOC through acid-catalyzed heterogeneous oxidation with hydrogen peroxide. In this study, we attempt to give a field evidence for the H2O2-WSOC relationship for the first time. Actually, our given evidence is still weak, and more field evidences should be provided in the future.

In addition to H2O2, organic hydroperoxides, especially HMHP and MHP, potentially play a role in the formation of WSOC. As mentioned above, HMHP can decompose to H2O2 and formaldehyde in the aqueous phase (O’Sullivan et al., 1996; Chen et al., 2008), which can subsequently participate in the formation of WSOC in the form of H2O2. It has been shown that the concentrations of H2O2 and MHP are similar in many parts of the atmosphere (Reeves and Penkett, 2003). Although the Henry’s Law constant of MHP in pure water is much lower than that of H2O2, the role of MHP in the atmospheric aqueous phase may be much more important than that estimated by its Henry’s Law constant in pure water. It is worth noting that formaldehyde was found in the aerosol at concentrations 1000-fold higher than the equilibrium concentration calculated only from its gas-phase formaldehyde and aqueous aerosol (Klippel and Warnack, 1978). This unexpected partitioning may be because formaldehyde in the aqueous aerosol is complexed with some soluble species (Facchini et al., 1992). The Henry’s Law constant of formaldehyde obtained in this case is usually called its effective Henry’s Law constant. However, to our knowledge, a similar study for enhanced solubility of MHP in the aqueous phase has not been reported. Considering the potential importance of MHP in the aqueous-phase reaction, its
effective Henry's Law constant in solutions regarding real atmospheric conditions needs further study.

In summary, hydroperoxides play an important role substantially in the formation of secondary sulfate and potentially in the formation of organic aerosols. First, hydroperoxides oxidize SO$_2$ into sulfate aerosols and simultaneously produce hydrogen ions. Second, with the increase of hydrogen ions derived from the above reaction, hydroperoxides will effectively oxidize organic compounds into WSOC by acid-catalyzed heterogeneous reactions. Third, the formation of WSOC will increase the hygroscopicity of aerosols, which in turn results in an increase of SO$_2$ oxidation by increasing the aqueous phase. Therefore, hydroperoxides serve as an important link between sulfate and organic aerosols. Such a link needs further study and should be considered in current atmospheric models.

4 Conclusions

Atmospheric H$_2$O$_2$ and organic hydroperoxides were measured for 13 days during the PRIDE-PRD’06 campaign at Backgarden, a rural site located ∼60 km northwest of Guangzhou. H$_2$O$_2$ and MHP were the dominant hydroperoxides present in the ambient air with a maximum mixing ratio of 4.6 ppbv for H$_2$O$_2$ and 0.8 ppbv for MHP. BHMP, PAA, HMHP, 1-HEHP and EHP were detected occasionally. H$_2$O$_2$ exhibited the maximum mixing ratio mainly between 12:00 and 18:00 LT on sunny days and low values at night and in the morning. Sometimes a second peak was observed during the evening (20:00–02:00 LT), which might be produced by the ozonolysis of alkenes. The diurnal variation of MHP was generally consistent with that of H$_2$O$_2$ but less pronounced. The estimation for the H$_2$O$_2$ formation rate from HO$_2$ recombination indicates that in the morning most of the H$_2$O$_2$ was formed through local photochemical activity, and vertical mixing might be a source. It was noteworthy that high levels of hydroperoxides were found in polluted air with a high mixing ratio of VOC and CO. The high level of HO$_2$ radicals and the low level of NO detected simultaneously in this region in the day may effectively support the production of hydroperoxides. High concentrations of H$_2$O$_2$ and MHP were detected in samples of rain collected during a shower when a strong typhoon passed through the observation site. The estimation using the Henry’s Law indicates that a considerably high mixing ratio of MHP resided above the boundary layer and might further influence the redistribution of HO$_X$ and RO$_X$ radicals in the PRD region. Evidence was found that hydroperoxides, in particular H$_2$O$_2$, contributed considerably to the formation of aerosol-phase sulfate via the aqueous-phase oxidation, and heterogeneous reactions may contribute substantially to the concentration of sulfate measured at the site. Furthermore, the results seem to imply that hydroperoxides may contribute potentially to the formation of WSOC, as indicated by the fact that their diurnal variations exhibited a negative correlation. This possibly provides evidence gathered in the field to support the importance of hydroperoxides in the formation of SOA found in laboratory studies. We suggest that hydroperoxides serve as an important link between sulfate and organic aerosols. This link needs further study and should be considered in current atmospheric models.

Acknowledgements. The authors gratefully thank the National Natural Science Foundation of China (grants 20677002 and 20107001), and the Project of Development Plan of the State Key Fundamental Research of MOST of China (grant 2002CB410801), for their financial support. The authors would like to thank F. Yang, S. Guo, and X. Li for their O$_3$, SO$_2$, CO and PM$_{2.5}$ measurements and data support (Peking University); H. Su for gas-phase HNO$_3$ measurement (Peking University); B. Wang for PAN measurement (Peking University); S.-J. Fan for meteorological data (Sun Yet-Sen University); and C.-H. Lai for VOC measurement (Research Center of Environment Changes, Academia Sinica, Taiwan).

Edited by: L. Carpenter

References

Ammann, M. and Pöschl, U.: Kinetic model framework for aerosol and cloud surface chemistry and gas-particle interactions – Part 2: exemplary practical applications and numerical simulations, Atmos. Chem. Phys., 7, 6025–6045, 2007, http://www.atmos-chem-phys.net/7/6025/2007/.

Andreae, M. O. and Crutzen, P. J.: Atmospheric aerosols: biogeochemical sources and role in atmospheric chemistry, Science, 276, 1052–1058, 1997.

Arellanes, C., Paulson, S. E., Fine, P. M., and Sioutas, C.: Exceeding of Henry’s law by hydrogen peroxide associated with urban aerosols, Environ. Sci. Technol., 40, 4859–4866, 2006.

Atkinson, R. and Aschmann, S. M.: OH production from the gas-phase reactions of O$_3$ with a series of alkenes under atmospheric conditions, Environ. Sci. Technol., 27(7), 1357–1363, 1993.

Atkinson, R.: Gas-phase tropospheric chemistry of organic compounds, J. Phys. Chem. Ref., Data, Monogr., 2, 1–216, 1994.

Becker, K.-H., Bechara, J., and Brockmann, K. J.: Studies on the formation of H$_2$O$_2$ in the ozonolysis of alkenes, Atmos. Environ., Part A, 27, 57–61, 1993.

Becker, K.-H., Brockmann, K. J., and Bechara, J.: Production of hydrogen peroxide in forest air by reaction of ozone with terpenes, Nature, 346, 256–258, 1990.

Böge, O., Miao, Y., Plewka, A., and Herrmann, H.: Formation of secondary organic particulate phase compounds from isoprene gas-phase oxidation products: an aerosol chamber and field study, Atmos. Environ., 40, 2501–2509, 2006.

Buffalini, J. J., Gay, B. W., and Brubaker, K. L.: Hydrogen peroxide formation from formaldehyde photooxidation and its presence in urban atmospheres, Environ. Sci. Technol., 6, 616–622, 1972.

Calvert, J. G., Lazrus, A. L., Kok, G. L., Heikes, B. G., Welega, J. G., Lind, J., and Cantrell, C. A.: Chemical mechanism of acid generation in the troposphere, Nature, 317, 27–35, 1985.

Chen, Z. M., Wang, H. L., Zhu, L. H., Wang, C. X., Jie, C. Y., and Hua, W.: Aqueous-phase ozonolysis of methacrolein and methyl...
vinyl ketone: a potentially important source of atmospheric aqueous oxidants, Atmos. Chem. Phys., 8, 2255–2265, 2008, http://www.atmos-chem-phys.net/8/2255/2008/.

Chin, M., Jacob, D. J., Munger, J. W., Parrish, D. D., and Doddridge, B. G.: Relationship of ozone and carbon monoxide over North America, J. Geophys. Res., 99, 14,565–14,573, 1994.

Chung, M. Y., Muthana, S., Paluyo, R. N., and Hasson, A. S.: Measurements of effective Henry’s law constants for hydrogen peroxide in concentrated salt solutions, 39, 2981–2989, 2005.

Claeys, M., Graham, B., Vas, G., Wang, W., Vermeulen, R., Pashynska, V., Cafmeyer, J., Guyon, P., Andreae, M.O., Artaxo, P., and Maenhaut, W.: Formation of secondary organic aerosols through photooxidation of isoprene, Science, 303, 1173–1176, 2004a.

Claeys, M., Wang, W., Ion, A. C., Kourtchev, I., Gelencsér, A., and Maenhaut, W.: Formation of secondary organic aerosols from isoprene and its gas-phase oxidation products through reaction with hydrogen peroxide, Atmos. Environ., 38, 4093–4098, 2004b.

Cohan, D. S., Schultz, M. G., Jacob, D. J., Heikes, B. G., and Blake, D. R.: Convective injection and photochemical decay of peroxides in the upper troposphere: methyl iodide as a tracer of marine convection, J. Geophys. Res., 104(D5), 5717–5724, 1999.

Das, M. and Aneja, V. P.: Measurements and analysis of concentrations of gaseous hydrogen peroxide and related species in the rural central Piedmont region of North Carolina, Atmos. Environ., 28, 2473–2483, 1994.

Facchini, M. C.,uzzi, S., Lind, J. A., Fierlingeroberlinninger, H., Kalina, M., Puxbaum, H., Winiwarter, W., Arends, B. G., Wobrock, W., Jaeschke, W., Berner, A., and Kruisz, C.: Phase-partitioning and chemical-reactions of lowmolecular-weight organic-compounds in fog, Tellus, 44, 533–544, 1992.

Fels, M. and Junkermann, W.: The occurrence of organic peroxides in air at a mountain site, Geophys. Res. Lett., 21, 341–344, 1994.

Finlayson-Pitts, B. J. and Pitts, J. N.: Atmospheric Chemistry: Fundamentals and Experimental Techniques, Wiley Interscience, John Wiley and Sons, New York, 1986.

François, S., Sowka, I., Monod, A., Tenime-Roussel, B., Laugier, J. M., and Worthing, H.: Development of an online analyzer of atmospheric H2O2 and several organic hydroperoxides for field campaigns, Atmos. Res., 74, 525–545, 2005.

Fung, C. S., Misra, P. K., Bloxam, R., and Wong, S.: A numerical experiment on the relative importance of H2O2 and O3 in aqueous conversion of SO2 to SO42−, Atmos. Environ., 25A(2), 411–423, 1991.

Gäb, S., Hellpolntner, E., Turner, W. V., and Korte, F.: Hydroxyethyl hydroperoxide and bis(hydroxyethyl) peroxide from gas-phase ozonolysis of naturally occurring alkenes, Nature, 316, 535–536, 1985.

Gäb, S., Turner, W. V., Wolff, S., Becker, K. H., Ruppert, L., and Brockmann, K. J.: Formation of alkyl and hydroxyalkyl hydroperoxides in ozonolysis in water and in air, Atmos. Environ., 29, 2401–2407, 1995.

Grossmann, D., Moortgat, G. K., Kibler, M., Kibler, M., Schlomski, S., Bachmann, K., Al intimidate, B., Geyer, A., Platt, U., Hamer, M. U., Vogel, B., Milacetic, D., Hofzumahaus, A., Holland, F., and Volz-Thomas, A.: Hydrogen peroxide, organic peroxides, carbonyl compounds, and organic acids measured at Pabstburg during BERLIOZ, J. Geophys. Res., 108(D4), 8250, doi:10.1029/2001JD001096, 2003.

Gunz, D. W. and Hoffmann, M. R.: Atmospheric chemistry of peroxides: a review, Atmos. Environ., 24, 1601–1633, 1990.

Hasson, A. S. and Paulson, S. E.: An investigation of the relationship between gas-phase and aerosol-borne hydroperoxides in urban air, J. Aerosol. Sci., 34, 459–468, 2003.

Hatakeyama, S. and Akimoto, H.: Reactions of creatie intermediates in the gas phase, Res. Chem. Intermed., 20, 503–524, 1994.

Heikes, B. G., Lee, M., Bradshaw, J., Sandholm, S., Davis, D. D., Chameides, W., Rodriguez, H., Liu, S., and McKeen, S.: Hydrogen peroxide and methyl hydroperoxide distributions over the North Pacific in the fall of 1991, J. Geophys. Res., 101, 1891–1905, 1996.

Hellpolntner, E. and Gäb, S.: Detection of methyl, hydroxymethyl and hydroxyethyl hydroperoxides in air and precipitation, Nature, 337, 631–634, 1989.

Hewitt, C. N. and Kok, G. L.: Formation and occurrence of organic hydroperoxides in the troposphere: Laboratory and field observations, J. Atmos. Chem., 12, 181–194, 1991.

Hoffmann, M. R. and Edwards, J. O.: Kinetics of oxidation of sulfate by hydrogen-peroxide in acidic solution, J. Phys. Chem., 79, 2096–2098, 1975.

Holland, F., Hofzumahaus, A., Schaefer, J., Kraus, A., and Paetz, H.-W.: Measurements of OH and HO2 radical concentrations and photolysis frequencies during BERLIOZ, J. Geophys. Res., 108(D4), 8246, doi:10.1029/2001JD001393, 2003.

Horie, O., Neeb, P., Limbach, S., and Moortgatt, G. K.: Formation of formic acid and organic peroxides in the ozonolysis of ethene with added water vapour, Geophys. Res. Lett., 21, 1523–1526, 1994.

Inuma, Y., Böge, O., Gnaeu, T., and Herrmann, H.: Aerosol-chamber study of the α-pinene/O3 reaction: influence of particle acidity on aerosol yields and products, Atmos. Environ., 38, 761–773, 2004.

Jackson, A. V. and Hewitt, C. N.: Hydrogen peroxide and organic hydroperoxide mixing ratios in air in a eucalyptus forest in central Portugal, Atmos. Environ., 30, 819–830, 1996.

Jacob, D. J., Heikes, B. G., Fan, S.-M., Logan, J. A., Mauzerall, D. L., Bradshaw, J. D., Singh, H. B., Gregory, G. L., Talbot, R. W., Blake, D. R., and Sachse, G. W.: The origin of ozone and NOx in the tropical troposphere: A photochemical analysis of aircraft observations over the south Atlantic basin, J. Geophys. Res., 101, 24235–24250, 1996.

Jacob, P., Tavares, T. M., Rocha, V. C., and Klockow, D.: Atmospheric H2O2 field-measurements in a tropical environment-Bahia, Brazil, Atmos. Environ., 124A, 377–382, 1990.

Jaegle, L., Jacob, D. J., Wennberg, P. O., Spivakovksy, C. M., Hanisco, T. F., Lanzendorf, E. J., Hintsa, E. J., Fahey, D. W., Keim, E. R., Proftt, M. H., Atlas, E. L., Flocke, F., Schaufler, S., McElroy, C. T., Midwinter, C., Pfister, L., and Wilson, J. C.: Observed OH and HO2 in the upper troposphere suggest a major source from convective injection of peroxides, 24, 3181–3184, 1997.

Jang, M. S., Carroll, B., Chandramouli, B., and Kamens, R. M.: Particle growth by acid-catalyzed heterogeneous reactions of organic carbonyls on preexisting aerosols, Environ. Sci. Technol., 37, 3828–3837, 2003.

Jang, M. S., Czoschke, N. M., Lee, S., and Kamens, R. M.: Heterogeneous atmospheric aerosol production by acid-catalyzed particle-phase reactions, Science, 298, 814–817, 2002.
Jayne, J. T. and Davidovits, P.: Uptake of SO$_2$(g) by aqueous surfaces as a function of pH: the effect of chemical reaction at the interface, J. Phys. Chem., 94, 6041–6048, 1990.

Johnson, R. M. and Siddiqi, I. W.: The determination of organic peroxides, Pergamon, New York, 119 pp., 1970.

Kelly, T. J., Daum, P. H., and Schwartz, S. E.: Measurements of peroxides in cloudwater and rain, J. Geophys. Res., 90, 7861–7871, 1985.

Kim, Y. M., Lee, M., Chang, W., Lee, G., Kim, K. R., and Kato, S.: Atmospheric peroxides over the North Pacific during IOC 2002 shipboard experiment, Chemosphere, 69, 1638–1646, 2007.

Klippel, W. and Warneck, P.: Formaldehyde in rain water and on atmospheric aerosol, Geophys. Res. Lett., 5, 177–179, 1978.

Kok, G. L., McLaren, S. E., and Staffelbach, T. A.: HPLC determination of atmospheric organic hydroperoxides, J. Atmos. Ocean. Tech., 12, 282–289, 1995.

Kroll, J. H., Ng, N. L., Murphy, S. M., Flagan, R. C., and Seinfeld, J. H.: Secondary organic aerosol formation from isoprene photooxidation, Environ. Sci. Technol., 40, 1869–1877, 2006.

Kurth, H.-H., Gáb, S., Turner, W. V., and Kettrup, A.: A high-performance liquid chromatography system with an immobilized enzyme reactor for detection of hydrophilic organic peroxydes, Anal. Chem., 63, 2586–2589, 1991.

Lazrus, A. L., Kok, G. L., Lind, J. A., Gitlin, S. N., Heikes, B. G., and Shetter, R. E.: Automated fluorimetric method for hydrogen peroxide in air, Anal. Chem., 58, 594–597, 1986.

Lee, J. H., Leahy, D. F., Tang, I. N., and Newman, L.: Measurement and speciation of gas phase peroxydes in the atmosphere, J. Geophys. Res., 98, P122–P130, 1993.

Lee, M., Heikes, B. G., and Jacob, D. J.: Enhancements of hydroperoxides and formaldehyde in biomass burning impacted air and their effect on atmospheric oxidant cycles, J. Geophys. Res., 103(D11), 13 201–13 212, 1998.

Lee, M., Heikes, B. G., and O’Sullivan, D.: Hydrogen peroxide and organic hydroperoxide in the troposphere: A review, Atmos. Environ., 34, 3475–3494, 2000.

Lee, M., Kim, J.-A., Kim, Y.-M., and Lee, G.: Characteristics of atmospheric hydrogen peroxide variations in Seoul megacity during 2002–2004, Sci. Total. Environ., 393, 299–308, 2008.

Lee, M., Noone, B. C., O’Sullivan, D., and Heikes, B. G.: Method for the collection and HPLC analysis of hydrogen peroxide and C$_2$ and C$_3$ hydroperoxides in the atmosphere, J. Atmos. Ocean. Tech., 12, 1060–1070, 1995.

Li, G. Y., Tang, X. L., Li, X. H., Yi, F., Sheng, G. Y., and Fu, J. M.: Composition and mutagenicity of particle sized fraction from urban particulate matter in Guangzhou City, Acta Scientiae Circumstantiae, 25(3), 319–323, 2005.

Li, L., Chen, Z. M., Zhang, Y. H., Zhu, T., Ji, L., and Ding, J.: Kinetics and mechanism of heterogeneous oxidation of sulfur dioxide by ozone on surface of calcium carbonate, Atmos. Chem. Phys., 6, 2453–2464, 2006. http://www.atmos-chem-phys.net/6/2453/2006/.

Li, L., Chen, Z. M., Zhang, Y. H., Zhu, T., Li, S., Li, H. J., Zhu, L. H., and Xu, B. Y.: Heterogeneous oxidation of sulfur dioxide by ozone on the surface of sodium chloride and its mixtures with other components, J. Geophys. Res., 112, D18301, doi:10.1029/2006JD008207, 2007.

Li, S., Chen, Z. M., and Shi, F.: Determination of Henry’s Law constant for methyl hydroperoxide by long path FTIR, Prog. Nat. Sci., 14(8), 31–35, 2004.

Li, Y. S., Campana, M., Reimann, S., Schaub, D., Stemmler, K., Staehelin, J., and Peter, T.: Hydrocarbon concentrations at the Alpine mountain sites Jungfraujoch and Arosa, Atmos. Environ., 39, 1113–1127, 2005.

Lightfoot, P. D., Cox, R. A., Crowley, J. N., Destriau, M., Hayman, G. D., Jenkin, M. E., Moortgat, G. K., and Zabel, F.: Organic peroxy radicals: kinetics, spectroscopy and tropospheric chemistry, Atmos. Environ., 26(10), 1805–1961, 1992.

Lind, J. A. and Kok, G. L.: Correction to ‘Henry’s Law determination for aqueous solutions of hydrogen peroxide, methyl hydroperoxide and peroxyacetic acid’, J. Geophys. Res., 99, 21119–21119, 1994.

Lind, J. A. and Kok, G. L.: Henry’s Law determination for aqueous solutions of hydrogen peroxide, methyl hydroperoxide and peroxyacetic acid, J. Geophys. Res., 91, 7889–7895, 1986.

Lind, J. A., Lazrus, A. L., and Kok, G. L.: Aqueous phase oxidation of sulfur (IV) by hydrogen peroxide and methyl hydroperoxide and peroxyacetic acid, J. Geophys. Res., 92, 4171–4177, 1987.

Lohmann, U. and Feichter, J.: Global indirect aerosol effects: a review, Atmos. Chem. Phys., 5, 715–737, 2005.

Madronich, S. and Calvert, J. G.: Permutation reactions of organic peroxyl radicals in the troposphere, J. Geophys. Res., 95(D5), 5697–5715, 1990.

Mair, R. D. and Hall, R. T.: Determination of organic peroxydes by physical, chemical, and colorimetric methods, Organic Peroxides Vol. 2, edited by: Swern, D., Wiley-Interscience, 535–635, 1970.

Mari, C., Jacob, D. J., and Bechtold, P.: Transport and scavenging of soluble gases in a deep convective cloud, J. Geophys. Res., 105(D17), 22 255–22 267, 2000.

Miyazaki, Y., Kondo, Y., Takegawa, N., Komazaki, Y., Fukuda, M., Kawamura, K., Mochida, M., Okuzawa, K., and Weber, R. J.: Time-resolved measurements of water-soluble organic carbon in Tokyo, J. Geophys. Res., 111, D23206, doi:10.1029/2006JD007125, 2006.

Moortgat, G. K., Grossmann, D., Boddenaerg, A., Dallmann, G., Ligon, A. P., Turner, W. V., Gäb, S., Slemer, F., Wiprecht, W., Acker, K., Kibler, M., Schlomski, S., and Bächmann, K.: Hydrogen peroxide, organic peroxide and carbonyl compounds determined during the BERLIOZ Campaign, J. Atmos. Chem., 42, 443–463, 2002.

Morgan, R. B. and Jackson, A. V.: Measurements of gas-phase hydrogen peroxide and methyl hydroperoxide in the coastal environment during the PARFORCE project, J. Geophys. Res.-Atmos., 107, 8109, doi:10.1029/2000JD000257, 2002.

Neeb, P., Sauer, F., Horie, O., and Moortgat, G. K.: Formation of hydroxymethylhydroperoxide and formic acid in alkene ozonolysis in the presence of water vapour, Atmos. Environ., 31, 1417–1423, 1997.

Novakov, T. and Penner, J. E.: Large contribution of organic aerosols to cloud-condensation-nuclei concentrations, Nature, 365, 823–826, 1993.

O’Sullivan, D. W., Heikes, B. G., Lee, M., Chang, C., Gregory, G., Blake, D., and Sachase, G.: The distribution of hydrogen peroxide and methyl hydroperoxide in the Pacific and South Atlantic, J. Geophys. Res., 104, 5635–5646, 1999.

O’Sullivan, D. W., Lee, M. Y., Noone, B. C., and Heikes, B. G.: Henry’s law constant determinations for hydrogen peroxide, methyl hydroperoxide, hydroxymethyl hydroperoxide, ethyl...
hydroperoxide, and peroxyacetic acid, J. Phys. Chem., 100(8), 3241–3247, 1996.
Paulson, S. E. and Orlando, J. J.: The reactions of ozone with alkene: An important source of HOx in the boundary layer, Geophys. Res. Lett., 23, 3727–3730, 1996.
Pena, R. M., Garcia, S., Herrero, C., and Lucas, T.: Measurements and analysis of hydrogen peroxide rainwater levels in a Northwest region of Spain, Atmos. Environ., 35, 209–219, 2001.
Penkett, S. A., Jones, B. M. R., Brice, K. A., and Eggleton, A. E. J.: The importance of atmospheric ozone and hydrogen peroxyde in oxidising sulphur dioxide in cloud and rainwater, Atmos. Environ., 13, 123–137, 1979.
Penner, J. E., Andreae, M., Annegarn, H., Barrie, L., Feichter, J., Hegg, D., Jayaraman, A., Leaitch, R., Murphy, D., Nganga, J., and Pitari, G.: Climate Change 2001: the Scientific Basis. Contribution of Working Group I to the Third Assessment Report of the Intergovernmental Panel on Climate Change, Cambridge University Press, 289–348, 2001.
Perry, K. D., Cahill, T. A., Schnell, R. C., and Harris, J. M.: Long-Range Transport of Anthropogenic Aerosols to the National Oceanic and Atmospheric Administration Baseline Station at Mauna Loa Observatory, Hawaii, J. Geophys. Res., 104(D15), 18 521–18 535, 1999.
Pöschl, U., Rudich, Y., and Ammann, M.: Kinetic model framework for aerosol and cloud surface chemistry and gas-particle interactions – Part 1: General equations, parameters, and terminology, Atmos. Chem. Phys., 7, 5989–6023, 2007, http://www.atmos-chem-phys.net/7/5989/2007/.
Ravetta, F., Jacob, D. J., Brune, W. H., Heikes, B. G., Anderson, B. E., Blake, D. R., Gregory, G. L., Sachse, G. W., Sandholm, S. T., Shetter, R. E., Singh, H. B., and Talbot, R. W.: Experimental evidence for the importance of convected methylhydroperoxide as a source of hydrogen oxide (HOx) radicals in the tropical upper troposphere, J. Geophys. Res., 106(D23), 32 709–32 716, 2001.
Ravishankara, A. R.: Heterogeneous and Multiphase Chemistry in the Troposphere, Science, 276, 1058–1065, doi:10.1126/science.276.5315.1058, 1997.
Reeves, C. E. and Penkett, S. A.: Measurements of peroxyde and what they tell us, Chem. Rev., 103, 5199–5218, 2003.
Rieche, A. and Hitz, F.: Uber monomethyl-hydroperoxide, Ber. Dtsch. Chem. Ges., 62, 2458–2472, 1929.
Rieche, A. and Meister: Uber peroxyde des formaldehyd: oxymethylhydroperoxyd, Ber. Dtsch. Chem. Ges., 68, 1468–1472, 1935.
Sander, S. P., Friedl, R. R., Ravishankara, A. R., Golden, D. M., Kolb, C. E., Kurylo, M. J., Huie, R. E., Orkin, V. L., Molina, M. J., Moortgat, G. K., and Finlayson-Pitts, B. J.: Chemical Kinetics and Photochemical Data for Use in Atmospheric Studies, JPL Publication 02-25, Evaluation No. 14, 5–35, 2003.
Sauer, F., Beck, J., Schustera, G., and Moortgat, G. K.: Hydrogen peroxide, organic peroxides and organic acids in a forested area during FIELDVOC94, Chemosphere, Global Change Sci., 3, 309–326, 2001.
Sauer, F., Limbach, S., and Moortgat, G. K.: Measurements of hydrogen peroxide and individual organic peroxides in the marine troposphere, Atmos. Environ., 31, 1173–1184, 1997.
Sauer, F., Schäfer, C., Neeb, P., Horie, O., and Moortgat, G. K.: Formation of hydrogen peroxide in the ozonolysis of simple alkene under humid conditions, Atmos. Environ., 33, 229–241, 1999.
Sauer, F., Schuster, G., Schäfer, C., and Moortgat, G. K.: Determination of H2O2 and organic peroxyde in cloud- and rain-water on the Kleiner Feldberg during FELDEX, Geophys. Res. Lett., 23, 2605–2608, 1996.
Saxena, P. and Hildemann, L.M.: Water-soluble organics in atmospheric particles: a critical review of the literature and application of thermodynamics to identify candidate compounds, J. Atmos. Chem., 24, 57–109, 1996.
Schwartz, S. E.: Gas- and aequous-phase chemistry of HO2 in liquid water clouds, J. Geophys. Res., 89, 11 589–11 598, 1984.
Seinfeld, J. H. and Pandis, S. N.: : From Air Pollution to Climate Change, Atmospheric Chemistry and Physics, Wiley, New York, 724–743, 1998.
Seuwen, R., and Warneck, K.: Oxidation of toluene in NOx free air: Product distribution and mechanism, Int. J. Chem. Kin., 28, 315–322, 1995.
Staffelbach, T. A., Kok, G. L., Heikes, B. G., McCully, B., Mackay, G. I., Karecki, D.R., and Schiff, H. I.: Comparison of hydroperoxide measurements made during the Mauna Loa Observatory photochemistry experiment 2, J. Geophys. Res., 101, 14 729–14 739, 1996.
Stockwell, W. R.: On the HO2 + HO2 reaction: Its misapplication in atmospheric chemistry models, J. Geophys. Res., 100, 11 695–11 698, 1995.
Su, F., Calvert, J. G., and Shaw, J. H.: Mechanism of the photooxidation of gaseous formaldehyde, J. Phys. Chem., 83, 3185–3191, 1979.
Takegawa, N., Miyaizaki, Y., Kondo, Y., Komazaki, Y., Miyakawa, T., Jimenez, J. L., Jayne, J. T., Worsnop, D. R., Allan, J., and Weber, R. J.: Characterization of an Aerodynamic Aerosol Mass Spectrometer (AMS): Intercomparison with other aerosol instruments, Aerosol Sci. Technol., 39, 760–770, 2005.
Takegawa, N., Miyakawa, T., Kondo, Y., Jimenez, J. L., Zhang, Q., Worsnop, D. R., and Fukuda, M.: Seasonal and diurnal variations of submicron organic aerosol in Tokyo observed using the Aerodynamic aerosol mass spectrometer, J. Geophys. Res., 111, D11206, doi:10.1029/2005JD006515, 2006.
Thompson, A. M.: The Oxidizing Capacity of the Earth’s Atmosphere: Probable Past and Future Changes, Science, 256(5060), 1157–1165, 1992.
U.S. Environmental Protection Agency: National Air Quality and Emissions Trends Report, 1999, p. 237, US Environmental Protection Agency, Office of Air Quality Planning and Standards, Research Triangle Park, NC, 2001.
Vácha, R., Slavíček, P., Mucha, M., Finlayson-Pitts, B. J., and Jungwirth, P.: Adsorption of Atmospherically Relevant Gases at the Air/Water Interface: Free Energy Profiles of Aqueous Solvation of N2, O2, O3, OH, H2O2, HO2, and H2O2, J. Phys. Chem., 108, 11 573–11 579, 2004.
Valverde-Canossa, J., Grossmann, D., Neeb, P., and Moortgat, G. K.: Ozonolysis of biogenic and anthropogenic alkene as a relevant source of tropospheric H2O2 and organic hydroperoxyde, in: Proceedings of the Eurotrac-2 Symposium 2000 on “Transport and Chemical Transformation in the Troposphere”, 27–31 March 2000, Garmisch-Partenkirchen, Germany, edited by: Midgley, P. M., Reuther, M., and Williams, M., Springer-Verlag, New York, (extended abstract on CD-ROM), 2001.
Valverde-Canossa, J.: Sources and sinks of organic peroxyde in the planetary boundary layer, Ph.D. Dissertation, Johannes
Guteneck-Univ., Mainz, 2004.

Walcek, C. J.: A theoretical estimate of O₃ and H₂O₂ dry deposition over Northeastern United States, Atmos. Environ., 21, 2649–2659, 1987.

Walker, S. J., Evans, M. J., Jackson, A. V., Steinbacher, M., Zellweger, C., and McQuaid, J. B.: Processes controlling the concentration of hydroperoxides at Jungfraujoch Observatory, Switzerland, Atmos. Chem. Phys., 6, 5525–5536, 2006, http://www.atmos-chem-phys.net/6/5525/2006/.

Wang, C. X. and Chen, Z. M.: Effect of CH₃OOH on the atmospheric concentration of OH radicals, Prog. Nat. Sci., 16 (11), 1141–1149, 2006.

Wang, J. L., Din, G. Z., and Chan, C. C.: Validation of a laboratory-constructed automated gas chromatograph for the measurement of ozone precursors through comparison with a commercial analogy, J. Chromatogr. A, 1027(1–2), 11–18, 2004.

Wang, T., Poon, C. N., Kwok, Y. H., and Li, Y. S.: Characterizing the temporal variability and emission patterns of pollution plumes in the Pearl River Delta of China, Atmos. Environ., 37(25), 3539–3550, 2003.

Watkins, B. A., Parrish, D. D., Buhr, S., Norton, R. B., Trainer, M., Yee, J. E., and Fehsenfeld, F. C.: Factors influencing the mixing ratio of gas phase hydrogen peroxide during the summer at Kinterbish, Alabama, J. Geophys. Res., 100, 22 841–22 851, 1995a.

Watkins, B. A., Parrish, D. D., Trainer, M., Norton, R. B., Yee, J. E., Fehsenfeld, F. C., and Heikes, B. G.: Factors influencing the mixing ratio of gas phase hydrogen peroxide during the summer at Niwot Ridge, Colorado, J. Geophys. Res., 100, 22 831–22 840, 1995b.

Wennberg, P. O., Hanisco, T. F., Jaegle, L., Jacob, D. J., Hintsa, E. J., Lanzendorf, E. J., Anderson, J. G., Gao, R. S., Keim, E. R., Donnelly, S. G., Del Negro, L. A., Fahey, D. W., McKeen, S. A., Salawitch, R. J., Webster, C. R., May, R. D., Herman, R. L., Proffitt, M. H., Margitan, J. J., Atlas, E. L., Schauffler, S. M., Flocke, F., McElroy, C. T., and Bui, T. P.: Hydrogen radicals, nitrogen radicals, and the production of O₃ in the upper troposphere, Science, 279(5347), 49–53, 1998.

Wesley, M. L.: Parameterization of surface resistances to gaseous dry deposition in regional scale numerical models, Atmos. Environ., 23, 1293–1304, 1989.

Xu, J. R. and Chen, Z. M.: Determination of peroxydes in environmental samples by high performance liquid chromatography with fluorescence detection, Chinese J. Chromatogr., 23, 366–369, 2005.

Zhang, Y. H., Shao, K. S., and Tang, X. Y.: The study of urban photochemical smog pollution in China, Acta Scientiarum Naturalium Universitatis Pekinesis, 24(2–3), 392–400, 1998.

Zhou, X. L. and Lee, Y.-N.: Aqueous solubility and reaction kinetics of hydroxymethyl hydroperoxide, J. Phys. Chem., 96, 265–272, 1992.