The dynamics of pine marten predation on red and grey squirrels

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Abstract
Invasive alien species pose one of the greatest threats to global biodiversity. In parts of Europe, introduced eastern grey squirrels (Sciurus carolinensis) have caused regional extinctions of the native red squirrel (Sciurus vulgaris). However, exposure to pine martens (Martes martes) has been demonstrated to reverse the competitive outcome between red and grey squirrels. The mechanism whereby this effect occurs remains unclear. It is hypothesised that direct predation, facilitated by a lack of behavioural response, is the mechanism driving this relationship. We review the literature and reanalyse a new dataset to provide further data on the occurrence of both squirrel species in the scats of pine marten. Both squirrel species occurred in the scats of pine marten confirming its role as a predator of these species. Predation of grey squirrels was significantly higher than red squirrels and was recorded only in spring and summer. Our review provides evidence for the mechanism driving the current decline in grey squirrels in Ireland and Scotland and supports the hypothesis that in the presence of a shared predator, direct predation influences the outcome of species interactions between native red and non-native grey squirrels.

Keywords Pine marten · Diet · Grey squirrel · Red squirrel · Predation · Invasive species · Prey switching

Introduction
Invasive alien species (IAS) arise when non-native species become established in a new environment and negatively impact the naive ecosystem. Impacts by IAS on native biodiversity, to which IAS are a global threat (ISAC 2006), have been widely reported. Awareness, and the ability to measure impacts, of IAS on economics, health and ecosystem services is increasingly being recognised and documented (Charles and Dukes 2008; Kelly et al. 2013; Pimental et al. 2000). The annual economic cost of IAS to Europe in 2009 was estimated at €12 billion p. a. (Kettunen et al. 2009), a cost that has likely increased due to continued increases of many IAS. Mitigation of these losses can take the form of direct culling (e.g. Schuchert et al. 2014), the use of poisons (Witmer et al. 2007) or biocontrol via the introduction of a predator (Lindquist et al. 1992) or a pathogen (Cooke and Fenner 2002). The introduction of disease or predators from the native range of an IAS (Simberloff and Stiling 1996; Shine 2010) may lead to unintended negative consequences, e.g. the development of resistance or predation on non-target species (Hays and Conant 2007; O’Donnell et al. 2017).

One IAS of widespread concern in Europe is the eastern grey squirrel (Sciurus carolinensis). Introduced from North America to Great Britain and Ireland in the nineteenth and twentieth centuries (Lloyd 1968), it is now in Italy where its presence poses a risk to the rest of continental Europe (Bertolino et al. 2014). The species negatively impacts forestry (Mayle and Broome 2013) and native biodiversity (Hewson and Fuller 2003; Newson et al. 2010), particularly the red squirrel which has been replaced across much of its former range in Britain (Lloyd 1968). Replacement of red by grey squirrels is due to factors including: larger individual mass and greater population densities of grey squirrels; decreases in red squirrel fecundity, residency and recruitment (Wauters and Gurnell 1999; Gurnell et al. 2004). Recent research suggests that alterations to microparasite communities (Romeo ...
et al. 2015) and personality (Wauters et al. 2019) in red squirrels during co-occurrence may also play a role. However, the most important factor remains grey squirrels acting as a vector for squirrel pox virus (SQPV) causing mortality in red, but not grey squirrels (Tomkins et al. 2003; Rushton et al. 2006). In order to prevent a repeat of this situation in Italy and continental Europe, efforts are required to manage or eradicate grey squirrels [Regulation (EU) 1143/2014].

Control of the grey squirrel has been attempted in Britain and Ireland via culling to mitigate their impacts. However, success has been localised and arguably ineffective, with few successful outcomes (Schuchert et al. 2014). Cost and effort required impedes successful control both locally and nationally in Great Britain and Ireland (Manchester and Bullock 2001), whilst in Italy, objections by animal rights groups have hindered control efforts (Genovesi and Bertolini 2001; Lioy et al. 2019). At present, culling is the only means available to control grey squirrels as efforts to develop an immuno-contraceptive for grey squirrels continue with no practical and deliverable method yet available (Barr et al. 2002). Consequently, a new approach is needed to provide effective control of the grey squirrel in European landscapes, reducing their economic cost (>£10 million per annum in Britain alone, Mayle and Broome 2013) and reversing their replacement of the native red squirrel.

The recent research suggests a potential new approach to control grey squirrels may involve a recovering native predator, the European pine marten (Martes martes). Negative, spatial correlations between pine marten and the presence of grey squirrels has been recorded in Ireland (Sheehy and Lawton 2014) and Scotland (Sheehy et al. 2018). The latter suggests exposure to pine martens negatively impacts grey, but not red squirrel occupancy (Sheehy et al. 2018) and that exposure to pine martens reverses the typical outcome of resource and disease mediated competition between the two sciurids. The recent research has demonstrated that native red squirrels possess anti-predator behaviours to pine martens, while invasive grey squirrels do not (Twining et al. 2020). This suggests behavioural differences may enable the red squirrel to avoid pine martens more successfully than greys which, as a result, are more vulnerable to predation by pine martens than red squirrels.

The European pine marten is an opportunistic omnivore which since its legal protection under the 1976 Irish Wildlife Act and the 1981 Wildlife and Countryside Act, has been recovering in the British Isles. Its recent recovery in Ireland and Britain is particularly interesting due to the fact the species is a woodland specialist, recovering in a landscape that is lacking in suitable habitat: woodland. The pine marten’s ability to do so has recently been linked to its dietary plasticity and proficiency for prey switching (Twining et al. 2019). Seasonal prey switching in response to abundance and vulnerability of resources may explain how pine marten recovery could have the observed effect of causing declines in grey squirrel populations (Sheehy and Lawton 2014; Sheehy et al. 2018; Flaherty and Lawton 2019), without squirrels being a frequently recorded item in the diet. Most dietary studies on pine marten either focus on single field sites or sampling periods (Helldin 2000; Lynch and McCann 2007; Zalewski 2007; Sheehy et al. 2013; O’Meara et al. 2014). The pine martens arboreality means it is the only mammalian predator in Ireland and Britain capable of targeting squirrels in their dreys. It may have a strong ability, therefore, to regulate prey populations lacking refugium from predation (Menge and Sutherland 1976).

In order to investigate the potential mechanism for the reversal of squirrel replacement, we reviewed published studies to compile available information on the occurrence of red and grey squirrels in the diet of the European pine marten, and reanalysed Twining et al. (2019), a new dataset of 918 scats. These were collected from twenty sites across the pine martens range in Northern Ireland accounting for the presence of red and grey squirrels at each site (Tosh 2015). We aim to determine whether there is evidence for pine marten predation to be a potential mechanism for the observed reversal of the replacement of red squirrels by grey squirrels in Britain and Ireland (Sheehy and Lawton 2014; Sheehy et al. 2018; Flaherty and Lawton 2019), and to highlight potential nuances in squirrel predation by the pine marten elsewhere in Europe.

**Methods**

Peer reviewed and unpublished literature on European pine marten diet was searched for via the Web of Science and Google Scholar. Searches conducted utilised the terms “pine marten diet squirrel”, “pine marten predation squirrel”. Publications were selected using the following criteria: (1) diet composition was estimated by analysis of scats with a minimum sample size of 100; (2) results were reported for each prey species as frequency of occurrence (FO %), or as number of times an item occurred and total sample size number to FO % could be calculated manually and (3) red or grey squirrels, or both species, were reported present in the study area or diet. In addition, we supplement the review material by using data from Twining et al. (2019) accounting for the presence of either squirrel species at the study sites. In the latter study, twenty sites were selected in which to collect scats and investigate pine marten diet (Twining et al. 2019) based on the data from a previous survey examining the distribution of the squirrel species and the pine marten in Northern Ireland (Tosh 2015). 3 km transects were established at each of the twenty study sites. All sites were visited once per month for 12 months and on the first visit, all scats were cleared. Subsequently, only fresh scats deemed
of the correct shape (e.g. heart or C-shaped) and smell (e.g. pungent, aromatic) typical of pine martens were collected (Lynch and McCann 2007; Twining et al. 2019). Scats were stored in sample bags with self-indicating silica gel to dry out samples. Developments in methods using mitochondrial DNA to assign species to scats (O’Reilly et al. 2007; Velli et al. 2019) has enabled some recent studies to make use of genetic techniques to determine species producing scats (Caryl et al. 2012). However, the current high costs of this method can be prohibitive to adoption of such an approach and many studies use alternative methods. Twining et al. (2019) used a multi-evidence approach, confirming the presence of pine martens at sites using camera traps followed by scat collection by a single trained fieldworker using a scoring system, with all scats in doubt being discarded. Further details and discussion of scat identification method is available in Twining et al. (2019).

Due to the scarcity of locations where pine marten and grey squirrel co-occur in Ireland (Sheehy and Lawton 2014; Tosh 2015; Flaherty and Lawton 2019), there is an unavoidable bias towards sites supporting only red squirrels. Sites included coniferous plantation, coniferous plantation and heath, deciduous forest, and mixed broadleaf and coniferous habitats.

Dry macro-analysis of scats was conducted to establish diet. Mammal hairs were identified to species level via examination of guard hair cuticle scales and medulla patterns at 10× and 40× magnification (Teerink 1991). Data are reported as frequency of occurrence (FO %) for comparison to previous dietary studies (Klare et al. 2011). Frequency of occurrence was calculated as

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\text{FO\%} = \frac{\text{Number of times prey item present}}{\text{Total number of scats}} \times 100
\]

Pearson’s Chi squared test was used to test for general effects of season on proportions of red and grey squirrel in the scats. Fisher’s Exact tests (FET) were used to compare proportions of grey and red squirrels in scats between seasons. Seasons were categorised using the meteorological Northern hemisphere definitions, i.e. Winter (Dec–Feb), Spring (Mar–May), Summer (June–August) and Autumn (Sept–Nov). All statistical tests were carried out in R. 3.2.1 (R Core Team 2014).

Results

The literature review revealed 13 publications (Table 1) in addition to Twining et al. (2019) that met the criteria above. Squirrel predation (grey or red) was recorded in 10 (76.9%) of the 13 publications (Table 1). Red squirrels were recorded as being present at the study sites of all 13 publications and occurred in the diet in 9 of these (69.2% of investigations where the species co-occurred with the pine marten). In contrast, grey squirrels were present in 1 of the 13 study sites as well as Twining et al. (2019), where they were also present in the diet (100% of investigations where the species co-occurred with the pine marten; Table 1). Overall, squirrel predation was highly variable, but typically low or absent for red squirrels with occurrence ranging from 0 to 19.0% (mean = 3.7%). The only study with occurrence of grey squirrels available to review had a significantly higher frequency of occurrence at 15.6% (Table 1). Despite low averages, there are rare examples of elevated predation rates on red squirrels, e.g. FO % = 50.1%, Storch et al. (1990); FO % = 11.5%, Pulliainen and Ollinmaki (1996). Both these studies were conducted during winters in boreal forests.

Year-round, seasonal data are scarce (Table 2). Only two published studies present data divided into four seasons, and neither study reports squirrel present in the diet (Sidorovich et al. 2005; Caryl et al. 2012). There is only one study with grey squirrels present in the studies reviewed, highlighting the scarcity of data on the occurrence of grey squirrels in pine marten diet when the species co-occur (Sheehy et al. 2013).

We re-examined data from Northern Ireland available from Twining et al. (2019). The investigation reported a total of 918 pine marten scats collected over 12 months, of which 763 were collected from the 15 red only sites, 48 from the two grey only sites and 107 from the two red and grey squirrel sites. These were red squirrel resident sites, but grey squirrels turned up in the scats, representing likely transient individuals. When we only considered pine marten scats from sites with either red or grey squirrels present (as confirmed by camera traps and presence of hair in the scats), grey squirrels had a mean frequency of occurrence of 12.0% (95% CI, 6.1–17.9%, n = 155), compared to 4.2% (95% CI 2.9–5.9%, n = 870) for red squirrels. Proportionally, grey squirrels occurred more often in pine marten scats than red squirrels (Fig. 1. Fisher Exact Test, p < 0.001).

The frequency of occurrence of both squirrel species in pine marten scats varied seasonally in the Northern Ireland samples (Fig. 2, grey squirrels: Pearson’s Chi squared, df = 3, \(X^2 = 13.663, p = 0.003\); red squirrels: Pearson’s Chi squared, \(df = 3, X^2 = 13.915, p = 0.009\)). Grey squirrels only occurred in pine marten scats during spring and summer, with no evidence of occurrence in autumn or winter. Red squirrels occurred in scats collected in spring, summer and also autumn. Like the grey squirrel, red squirrels did not occur in scats during winter. Frequency of occurrence of grey squirrel in scats did not differ between spring and summer (Fisher Exact Test, \(p = 1\)). Similarly, there was no difference in the frequency of occurrence of red squirrels between spring and summer (Fisher Exact test, \(p = 0.458\)), or autumn (Fisher Exact Test, \(p = 0.663\)). When the frequency of occurrence...
of the two squirrel species was compared, grey squirrels occurred proportionally more than red squirrels in summer (Fisher Exact Test, \( p = 0.034 \)), but there was no significant difference in spring (Fisher Exact Test, \( p = 0.165 \)).

## Discussion

The present report sheds light on the direct interactions of a recovering native European predator, the pine marten on both an invasive species, the grey squirrel and a European native, the red squirrel. In response to the scarcity of records in the literature, we provide new evidence to support pine marten predation on both native and invasive sciurid species. There was a higher proportional occurrence of grey squirrel than red squirrel remains in scats. Thus, we hypothesise and provide evidence in support of a plausible mechanism for the observed decline of grey squirrels (Flaherty and Lawton 2019), whereby a recovering predator population has reversed the direction of the interaction between two congeneric squirrel species (Sheehy and Lawton 2014; Sheehy et al. 2018).

A review of the literature demonstrates that pine martens are opportunistic omnivores with carnivorous tendencies with neither squirrel species being a typically commonly occurring prey item (Helldin 2000; Lynch and McCann 2007; Zalewski 2007; Romaniuk 2018; Twining et al. 2019). Their diet and resource use is dependent on the prey availability and vulnerability (Twining et al. 2019). However, investigations into pine marten diet have typically shown red squirrels to be numerically unimportant food items with occurrence in the diet to be generally low (Helldin 2000; Lynch and McCann 2007; Zalewski 2007; O’Meara et al. 2014; Sheehy and Lawton 2014), or absent (Balharry unpublished, Caryl et al. 2012). However, occasional reports of elevated occurrence of red squirrels in the scats of pine martens have been associated with cyclical crashes in microtine populations and noted increases in red squirrel populations (Pulliainen and Ollinmaki 1996; Storch et al. 1990). These

### Table 1

| Country          | Habitat                                                                 | Season               | Sample size \((n)\) | Red squirrel occurrence \(\%\) | Grey squirrel occurrence \(\%\) | Authors                        |
|------------------|-------------------------------------------------------------------------|----------------------|----------------------|-------------------------------|-------------------------------|--------------------------------|
| Northern Ireland | Broadleaf, Coniferous plantation, Coniferous plantation and heath, Mixed woodlands | All year             | 918                  | 4.8                           | 12.0                          | Twining et al. (2019)          |
| Republic of Ireland | Broadleaf forest                                                        | All year             | 387                  | 0.4                           | n/a                           | Lynch and McCann (2007)        |
| Republic of Ireland | Unspecified fragmented forest                                           | Spring and Summer    | 110                  | 5                             | 15.6                          | Sheehy et al. (2013)           |
| Republic of Ireland | Unspecified fragmented forest                                           | Summer               | 252                  | 3.1                           | n/a                           | O’Meara et al. (2014)          |
| Scotland         | Coniferous plantation                                                   | All year             | 2475                 | 0                             | n/a                           | Caryl et al. (2012)            |
| Scotland         | Coniferous plantation                                                   | Spring and Summer    | 158                  | 0.6                           | n/a                           | Grabham et al. (2019)          |
| Scotland         | Heath, coniferous plantation and native woodland                        | All year             | 1304                 | 0                             | n/a                           | Balharry, Unpub. thesis        |
| Finland          | Boreal forest                                                           | Winter               | 2198                 | 4.4                           | n/a                           | Pulliainen and Ollinmaki (1996) |
| Poland           | Old growth deciduous                                                   | All year             | 1203                 | 2.5                           | n/a                           | Zalewski (2007)               |
| Poland           | Old growth deciduous                                                   | All year             | 1735                 | 3.7                           | n/a                           | Jedrzejewski et al. (1993)    |
| Sweden           | Boreal forest                                                           | Summer               | 144                  | 38.0                          | n/a                           | Storch et al. (1990)           |
| Sweden           | Boreal forest                                                           | Winter               | 144                  | 19.0                          | n/a                           | Storch et al. (1990)           |
| Sweden           | Boreal forest                                                           | Winter and Summer    | 760                  | 1.5                           | n/a                           | Helldin (2000)                |
| Belarus          | Mixed coniferous and deciduous                                          | All year             | 1222                 | 0                             | n/a                           | Sidorovich et al. (2005)      |
| Russia           | Unspecified                                                             | Winter and Summer    | 171                  | 1.7                           | n/a                           | Yazan (1970)                  |

All references that had data to report are highlighted bold.
Table 2  Table showing review of seasonal red and grey squirrel frequency of occurrence in the diet of European pine marten published literature throughout the range of the pine marten.

| Season                  | Country            | Red squirrel occurrence (%) | Grey squirrel occurrence (%) | Authors                           |
|-------------------------|--------------------|------------------------------|------------------------------|----------------------------------|
| Autumn                  | Scotland           | 0                            | n/a                          | Caryl et al. (2012)              |
| Autumn                  | Belarus            | 0                            | n/a                          | Sidorovich et al. (2005)         |
| Autumn                  | Northern Ireland   | 1.4                          | 0                            | Twining et al. (2019)            |
| Autumn and Winter       | Poland             | 5.0                          | n/a                          | Zalewski (2007)                 |
| Autumn and Winter       | Poland             | 7.1                          | n/a                          | Jedrzejewski et al. (1993)      |
| Spring                  | Scotland           | 0                            | n/a                          | Caryl et al. (2012)              |
| Spring                  | Poland             | 5.6                          | n/a                          | Zalewski, 2007                  |
| Spring                  | Belarus            | 0                            | n/a                          | Sidorovich et al. (2005)         |
| Spring                  | Northern Ireland   | 6.8                          | 14.3                         | Twining et al. (2019)            |
| Spring and Summer       | Scotland           | 0.6                          | n/a                          | Grabham et al. (2019)            |
| Spring and Summer       | Republic of Ireland| 5                            | 15.6                         | Sheehy et al. (2013)             |
| Spring and Summer       | Poland             | 0.8                          | n/a                          | Jedrzejewski et al. (1993)      |
| Summer                  | Republic of Ireland| 3.1                          | n/a                          | O’Meara et al. (2014)            |
| Summer                  | Scotland           | 0                            | n/a                          | Caryl et al. (2012)              |
| Summer                  | Poland             | 0.5                          | n/a                          | Zalewski (2007)                 |
| Summer                  | Sweden             | 0.6                          | n/a                          | Helldin (2000)                   |
| Summer                  | Belarus            | 0                            | n/a                          | Sidorovich et al. (2005)         |
| Summer                  | Northern Ireland   | 13.8                         | 22.8                         | Twining et al. 2019             |
| Winter                  | Finland            | 4.4                          | n/a                          | Pulliainen and Ollinmaki, 1996   |
| Winter                  | Scotland           | 0                            | n/a                          | Caryl et al. (2012)              |
| Winter                  | Belarus            | 0                            | n/a                          | Sidorovich et al. (2005)         |
| Winter                  | Northern Ireland   | 0                            | 0                            | Twining et al. (2019)            |
| Winter                  | Sweden             | 2.4                          | n/a                          | Helldin (2000)                   |
| Winter and Summer       | Sweden             | 19.0                         | n/a                          | Storch et al. (1990)             |

Frequency of occurrence has been split into specific seasons where data available
All references that had data to report are highlighted bold

Fig. 1  Mean (95% clm) occurrence of grey and red squirrel in scats of the European pine marten (Martes martes) at all sites for red or grey squirrels (Sciurus spp.) Data from Twining et al. (2019)
records, along with the elevated predation of red squirrels in two commercial plantation sites in Ireland observed in the data from Twining et al. (2019), suggest that red squirrels serve as an important resource either when highly abundant, or when other key resources are scarce. These observations align with observations that pine martens are highly opportunistic and will target resources that are locally abundant and or vulnerable (Twining et al. 2019).

Despite being a topical and critical question (Sheehy and Lawton 2014; Sheehy et al. 2018; Flaherty and Lawton 2019), there have been limited investigations into the diet of pine martens where they co-occur with grey squirrels. However, when they do co-occur, grey squirrels are a relatively common food item (Sheehy et al. 2013). The lack of studies with grey squirrels present in the environment is likely due to the fact pine martens and grey squirrel ranges do not naturally overlap, and instances in which these species co-occur are highly ephemeral (Flaherty and Lawton 2019). Due to Italy being the only country in mainland Europe where both squirrel species and pine martens occur, Italian dietary studies are of particular interest (Balestrieri et al. 2011; Lombardini et al. 2015). However, neither were used in this review as neither squirrel species is present in Sardinia (Lombardini et al. 2015) and, although the red squirrel was present in at the sites near Garzaia di Valenza, it was not recorded in the diet (Balestrieri et al. 2011). Grey squirrels did not occur at either Italian study site. There is a gap in the literature pertaining to locations where grey squirrels and pine martens co-occur.

We provide much needed data on the occurrence of grey squirrels in the diet within the range of the European pine marten and confirm that pine martens prey upon grey squirrels exclusively in spring and summer. This aligns with the reported bi-annual breeding patterns of grey squirrels with litters produced between March and April and a second litter in July and August (Webley and Johnson 1983; Gurnell 1983, 1987). This alignment of increased predation suggests that pine martens may be predating grey squirrels whilst juveniles are present and still restricted to dreys. If pine martens were to take both juveniles and adults, the effect of this predation would likely have a profound effect on recruitment reducing overall population size (Gervasi et al. 2012). Population changes in many mammal species are sensitive to changes in survival (Haydon et al. 1999), whether pine marten predation focusses on juveniles, adults or both, predation in spring and summer could impair recruitment and consequently affect population size. A future study which followed the dynamics of a grey squirrel population closely, combined with monitoring of squirrel boxes with camera traps, while looking at the occurrence of grey squirrels in the scats of the pine marten would provide useful additional insights. Although the age of squirrels predated by pine martens requires direct quantification to make definitive conclusions, a bias towards consumption of juvenile squirrels parallels observations that pine martens significantly increase predation on birds by focussing on altricial juvenile and fledgling birds in spring and summer (Lynch and McCann 2007; Twining et al. 2018). It is possible that predation of grey squirrels in dreys during spring and summer is a by-product of the observed switch to more arboreal foraging targeting bird nests observed during these months (Twining et al. 2019).

The overall frequency of occurrence of grey squirrel remains in pine marten scats in the present study was similar to Sheehy and Lawton (2014). The latter had reservations due to low sample size but sample size here (n = 155) is less of
a concern. Sample size of pine marten scats in red squirrel areas of occurrence was much higher \((n=877)\). This disparity reflects the particular biogeographical circumstances in Ireland where grey squirrels and pine marten, increasingly, do not co-occur (Sheehy and Lawton 2014; Flaherty and Lawton 2019). Despite having similar reproductive cycles to grey squirrels and sharing similar seasonal peaks of presence in the diet of pine martens in spring and summer, red squirrels have a lower overall frequency of occurrence in pine marten scats than grey squirrels. Interspecific disparities in frequency of occurrence in the diet of similar prey species, however, may reflect differences in prey density, body size and behaviour (Toscano et al. 2010; Lucas and Rosenheim 2011; Tucker and Rogers 2014). This prevents a definitive conclusion explaining the more frequent occurrence in pine marten scats of greys than red squirrels. Differences in species ecology, with red squirrels living at much lower densities than greys \((0.3–1.5/ha\) vs. \(2–16/ha\); Gurnell 1987), making them less available as a food item in the environment, has been suggested as a potential basis for higher predation rates amongst grey squirrels (Sheehy and Lawton 2014). The evidence that exposure to pine martens reverses the outcome of red-grey squirrel competition at landscape scales throughout Ireland and Scotland, where densities of species vary greatly, suggests that density of squirrel populations does not greatly affect the underlying mechanism of this process. The recent research demonstrating that grey squirrels do not possess anti-predator behaviours in response to the pine marten, while red squirrels show very clear behavioural responses and avoidance may provide explanation of the observed higher occurrence of grey squirrels in the diet of pine marten (Twining et al. 2020).

In conclusion, although red squirrel predation by the European pine marten is typically low, there are specific situations when higher predation levels occur. Additionally, where grey squirrels and pine martens co-occur, occurrence of grey squirrels in the diet of the pine marten is significantly higher than that of red squirrels. Thus, we hypothesise that, due to the absence of anti-predator behaviours (Twining et al. 2020), the invasive grey squirrel is predated by pine marten proportionally more than the red squirrel leading to a gradual decline in grey squirrel abundance and distribution. Although not conclusive, we provide a plausible mechanism for the established reversal of native red squirrel by invasive grey squirrels in the presence of pine martens in Ireland and Britain (Sheehy and Lawton 2014; Sheehy et al. 2018; Flaherty and Lawton 2019).

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