Full Length Research Paper

Energy consumption and expenditure of *Panthera pardus* in the Southern African region: How much hunting success is enough?

Ielyzaveta Maksymivna Ivanova

School of Animal, Plant and Environmental Sciences, Faculty of Science, University of the Witwatersrand, South Africa.

Received 1 September, 2017; Accepted 23 October, 2017

*Panthera pardus* is a widespread mammalian carnivore, with a very broad diet range. Therefore, it is often seen as needing less protection as compared with some other predator species. However, with a 37% reduction in historic range and some subspecies critically endangered, the leopard is a species that does indeed require certain conservation attention. In Southern Africa, there are several threats facing the leopard: habitat loss, poaching, as well as killings associated with leopard-human conflict (the latter being aggravated by poorly-stocked reserves where the animals reside). In setting aside formal protected areas for the leopard and its prey, it is important to plan and stock these reserves in such a manner so as to limit potential conflict with owners of surrounding farmlands or tribal land. Focusing on the average daily energetic consumption and expenditure of the leopard in Southern Africa, this paper seeks to determine how regular successful hunts can help maintain the animal. It was found from the study that there is a very close balance between the energy consumption and expenditure of the leopard. Depopulation of a varying intensity may result from a hunting success probability below 0.5. Leopards are unlikely to persist where hunting success is reduced to 0.1 due to prey shortage. This finding is believed to provide some preliminary guidance for leopard prey stocking rates in the future.

**Key words:** African leopard, energy expenditure, prey stocking rate, leopard-farmer conflict, conservation, predator-prey interaction.

INTRODUCTION

There are a wide suite of conservation issues in the modern age (Gurevitch and Padilla, 2004; Giam et al., 2010; Bellard et al., 2012), each having a variety of often negative implications on global biodiversity, and each demanding attention (Sheil, 2001), research, and funding to address (or at the very least, to attempt to do so). One of the biggest conservation concerns, and one that is very often implicated in extinctions of species in modern times, is habitat loss (Simberloff, 1984; Tilman et al., 1994). Habitat loss causes a reduction in population size,
leaving a species more prone to the effects of stochastic events (Burkey, 1995), and thus increasing the potentiality of its extinction. The loss of habitat in general tends to outweigh the effects of fragmentation of such habitat (Fahrig, 1997), and is aggravated by land conversion tending to be non-random (Seabloom et al., 2002), biased towards areas valued agriculturally or those with a particular geographic placement.

Panthera pardus is classified as a vulnerable predatory species according to the latest IUCN Red List (Stein et al., 2016). It has a wide distribution comprising a broad region in Africa (with the Sahara Desert excluded from its range), the Arabian Peninsula, southwest Asia, as well as a small population maintained in the Russian Far-East (Nowell and Jackson, 1996). While some estimates show its Southern African range as being in no immediate danger of a severe decline (Martin and de Meulenaer, 1988), there have been criticisms of these estimates (Norton, 1990), as there are marked reductions in areas of encroaching human settlement and other habitat conversions, with range decline for the species being as high as 37% in a period of 100 years (Ray et al., 2005).

P. p. pardus is an African sub-species of the leopard, with some 78% of the overall species range occupied by this variant (Jacobson et al., 2016). South Africa is responsible for some of the biggest habitat losses for the species, with leopards in unprotected areas being severely restricted in their occurrence (Skead, 2007). One case of this was noted in the Phinda-Mkhuzo Complex, a small region lying along the eastern edge of South Africa, where the observed number of leopards was on average 11.11/100 km² within the protected Mkhuzo Game Reserve (core protected area), 7.17/100 km² in the neighbouring Phinda Private Game Reserve (buffer protected area), and then only 2.49/100 km² in the surrounding non-protected lands used for livestock farming, private game ranches and tribal land (Balme et al., 2010). With poaching and trophy hunting still being prominent factors driving leopard population in South Africa, there is an added potential risk of leopards being subjected to retaliatory farmer killings because of the real and perceived threats that these animals pose to livestock, as leopards tend to easily cross boundary fences (Balme et al., 2009; Chapman and Balme, 2010). Therefore, despite about 20% of South Africa currently providing suitable leopard habitat (Swanepoel et al., 2013), there is a need for a thorough investigation of the potential of the species to experience a further decline, as suggested by the decreasing population trend predicted by the IUCN (Stein et al., 2016).

An important factor to consider when determining the potentiality of a population decline is to examine the species’ physiological needs against the resources that the environment in which the species occurs is actually capable of providing at any given time (Wikelski and Cooke, 2006). If the surrounding environment falls short of meeting an animal’s physiological needs, the animal is faced with a decision to either relocate elsewhere in search of resources or to face death from starvation in its current habitat (Stephens, 2008). However, where threats such as poaching and legal hunting outside the habitat pose a life-threatening risk of their own (as earlier mentioned) – the surrounding matrix is in itself an unpredictable habitat – the animal may still face the same fate outside the poor habitat as it would by remaining there (Switzer, 1993). Therefore, using the physiological needs of species to determine the required energetic quality of the habitat to maintain a viable population is a useful technique in helping inform habitat management (particularly where such habitat is facing the potentiality of, for example, habitat loss (Fahrig, 2001)). When assessing the energetic needs of a predator, it is of fundamental importance to consider the interactions that they have with their potential prey – a relationship that is instrumental in regulating and shaping both populations and communities at large (Fretwell, 1987). Not only does the predator-prey relationship affect prey numbers as per predator kills, but extensive phenotypic changes can be induced in the prey as a response (Werner and Peacer, 2003), and prey intimidation has an effect on their demographics comparable to those resulting from prey consumption (Preisser et al., 2005).

In accounting for the predator-prey interaction that exists between leopards and their prey (mostly mammalian species with a weight range of 10 to 40 kg, as per Hayward et al., 2006), it is also important to consider the effects of spatial heterogeneity on modifying the functional response (Gorini et al., 2012). In a heterogeneous system that tends to persist in real-life scenarios, the leopard will face regular challenges not only in searching for and encountering prey, but also in actually killing and consuming it.

A factor complicating the matter even further in the modern times is human disruption: anthropogenic activities can directly influence the nature of the relationship between a predator and its prey, whether by controlling the numbers of the respective species, or by providing food subsidies (Rodewald et al., 2011). In the case of the leopard, subsidies may be provided accidentally, such as domestic livestock that the predator may take to hunting (Kissui, 2008). Being assisted by the natural tendency of leopards to roam widely in search of prey, the animals can easily become involved in tense human-animal interactions with the owners of private lands (as earlier mentioned). As such, in assisting the survival of the species, conservation efforts need to be adequately backed by knowledge of how much wild prey should be stocked in reserves, to allow leopard densities to stabilize in accordance with their main natural prey, as is common for predators (Karanth et al., 2004). This should also assist in reducing incidences of livestock killing. To determine this, a detailed analysis is needed
of the difference between energetic consumption and expenditure of the animal, and how vulnerable it actually is to having its energy intake fall below its rate of utilization, upon which a steady decline in weight and activity levels ensues and the potential of death becomes a serious threat. This parameter is closely examined henceforth.

**METHODOLOGY**

An energetic model for *P. pardus* was constructed, using standard work (W) equations (presented in kilojoule (kJ) measurement units), which relates directly to kinetic energy according to the work-energy principle. Input values for the model were sourced from a variety of literature, obtaining specific values on feeding, hunting and other behaviours as is accurate for the species.

For the purposes of standardization, the *p. pardus* subspecies was chosen as the subject to allow for maximal consistency of the input values, given the well-defined genetic differences between the various subspecies (Miththapala et al., 1996). Where required energetic values were not available specifically for the African leopard, the closest approximation was chosen, that is, first looking for values concerning the leopard, followed by a search for felids at large, etc. The model environment selected for the subject to operate in was the savanna biome of Southern Africa, more specifically the northern Kwa-Zulu Natal (due to the concentration of leopard studies in that region), obtaining the physiologically relevant seasonal temperatures, the diurnal/nocturnal differences as appropriate for the region, as well as other atmospheric properties such as average wind speed, from literature sources.

Following this, a statistically-average African leopard was defined, to allow for the refinement of the energetic results obtained. An average daily energetic gain was determined for the leopard subject, using the values obtained for prey consumption. On the opposite end of the scales, the basal metabolic energy consumption of the subject was determined, followed by heat-related costs to the organism (such as due to radiation, convection, etc.). Lastly, energy costs associated with obtaining prey were calculated. Summing all the determined energetic costs and weighing those up against the average energetic gains the leopard can be expected to obtain, allowed the determination of the extent of the difference between the energy gains and losses. The energetic surplus would be used by the organism to meet other, non-daily, needs such as breeding effort or emigration out of an area (Parker et al., 2009).

The results of the model provide another important finding: an estimation of how many hunting attempts on average before success are possible to warrant an organism's continued survival. Using this value, various scenarios of catch probability were tested to determine how many hunting attempts were required before a catch was made under each case. Given that a higher prey density is strongly related to the incidence of a successful catch (Whitfield, 2003), catch probability served as a proxy for prey stocking rates in the environment tested. This allowed for the determination of the lowest hunting success probability (that is, prey stocking rates) at which the leopard would still be able to obtain a hunt before the average number of hunting attempts, as allowed by standard daily energy reserves, were exhausted.

**RESULTS**

One of the main factors involved in energy expenditure, is that required for the correct functioning of internal organs such as the heart, the lungs, liver, etc. Even while being within a state of relative physical inactivity, there is a continuous utilization of energy for the continuation of basal metabolism, characterized by catabolism of compounds with oxygen intake.

The rate of oxygen consumption by the animal, the energetic output produced, as well as the rate of carbon dioxide release, is interrelated. If carbohydrates are subjected to oxidation, the intake of one litre of oxygen frees up 21.13 kJ of energy. If there is a similar oxidation of proteins and lipids, then 19.66 and 20.08 kJ of energy are released, respectively. For an adult animal on average, the hourly basal metabolism uses up 4.2 kJ for a kilogram of body weight.

Energetic balance analysis defines the following parameters for an adult leopard of mean statistical parameters in Southern African savanna: a body mass of 60 kg and a body surface area of 1.53 m² (using Meeh coefficient of 10 for an average cat, given by Schmidt-Nielsen, 1984).

*P. pardus* has an internal body temperature of about 39.86°C (Deka et al., 2012), while the ambient temperature in its savanna habitat fluctuates seasonally (Balme et al., 2007). Midday temperatures range from 23°C in July to 30°C in January (BirdLife South Africa, 2016), while night-time temperatures of 11°C in July and 20°C in January. Averaging, we obtain: 26.5°C in summer and 15.5°C in winter. The epidermal layers of an animal tend to be a few degrees cooler than the rectal temperatures as recorded by Deka et al. (2012), with the epidermis being where internal and ambient temperatures meet (ambient temperatures averaging about 26°C in Southern African savanna). Given the aforementioned considerations, this study assumes an average epidermal temperature to be about 30°C for the African leopard (*p. pardus*).

Further, the study assumes that a statistically-average African leopard in African savanna has a daily meat consumption of 3.25 kg, since this is the median value of a leopard’s 1.6-4.9 kg meat/day consumption recorded in the literature (Bothma and Le Riche, 1986; Bailey, 1993; Stander et al., 1997). African leopards are known to strongly prefer killing impala and bushbuck, with an average body mass of 23 kg (Hayward et al., 2006). The energetic content of this game can be estimated at 8.5 kJ/g, since venison is considered to be a much leaner type of meat than that of cattle, with a common fat content being less than 3% (Schönfeldt, 1993; Hoffman, 2000). The quantity of energy obtained by an African leopard in a single statistically-average 24-h period therefore becomes:

$$W_{EO} = 8.5kJ \times (1000 \times 3.25kg) = 27625 \text{ kJ}.$$ Utilization of the energy source provides for sustenance
of a leopard’s life until the next successful hunt. Loss of energy due to basal metabolism can be presented as:

\[ W_{\text{BM}} = 4.2 \text{kJ} \cdot 24 \text{ h} \cdot 60 \text{ kg} = 6048 \text{ kJ} \]

Alongside the basal metabolism energetic expenditure, there are significant costs associated with heat exchange between the body and the exterior environment. In the complex process of maintenance of a heat balance, a major importance is the intricate regulation of heat loss (Berkovich, 1964). In physiology, bodily heat transfer can be viewed as the loss of heat, freed through activities associated with living, into a cooler environment (Ivanov, 1990). There are four key modes of heat transfer between an animal and its environment: radiation, convection, conduction, and evaporation; the latter being dominant in cases of overheating. However, when existing under conditions of a comfort temperature zone, the greatest exchange is provided by radiation and convection (Fanger, 1970).

The total energy radiated from a unit of bodily surface area is directly proportional to the fourth power of the thermodynamic absolute temperature, as detailed by the Stefan-Boltzmann law. When there is but a small difference between the animal's epidermal temperature and the ambient temperature (as is generally the case in the Southern African part of a leopard’s distribution), the equation for radiative heat loss can be presented in the following format:

\[ W_R = h_{\text{rad}} \cdot S \cdot (\delta_1 - \delta_2) \cdot t, \]

where \( h_{\text{rad}} \) is the radiative heat transfer coefficient of \( W_R \) (in \( \text{m}^2 \cdot \text{C}^{-4} \)), \( S \) is the leopard’s body surface area (in \( \text{m}^2 \)), \( \delta_1 \) is the leopard’s epidermal temperature, \( \delta_2 \) is the ambient temperature, and lastly, \( t \) is length of time over which radiation is being measured (in seconds).

By the Stefan-Boltzmann law, the radiative heat transfer coefficient \( h_{\text{rad}} \) between two gray surfaces can be determined with the equation:

\[ h_{\text{rad}} = \varepsilon \sigma (T_1^2 + T_2^2) / (T_1 + T_2), \]

where \( \varepsilon \) is the emissivity of the leopard's epidermis, \( \sigma \) is the Stefan-Boltzmann constant (\( \sigma = 5.67 \times 10^{-8} \text{ \text{W} \cdot \text{m}^{-2} \cdot \text{K}^{-4}} \)), \( T_1 \) is the epidermal temperature of the leopard and \( T_2 \) is the ambient temperature (both temperatures being absolute).

Emissivity varies with the radiation wavelength, but is close to unity at wavelengths greater than 5 \( \mu \text{m} \) (Ingram and Mount, 1975), therefore a value of 0.9 is here assumed for \( \varepsilon \). \( T_1 \) is 30°C and \( T_2 \) is 26.5°C for daytime, with \( T_2 \) becoming 15.5°C average at night (as explained previously). From the abovementioned, the calculation for \( h_{\text{rad}} \) becomes:

\[ (0.9) (5.67 \times 10^{-8}) (303^2 + 299.5^2) (303 + 299.5) = 5.58 \text{ during the day, and} (0.9)(5.67 \times 10^{-8})(303^2 + 288.5^2) (303 + 288.5) = 5.28 \text{ at night.} \]

Therefore for \( W_R \), we now have:

\[ W_R = 5.58 \times 1.53 \times (30 - 26.5) \times 3600 \text{ ss} \times 12 \text{ h} = 29.88 \text{ J/s} \times 3600 \text{ seconds} \times 12 \text{ h} = 1290.85 \text{ kJ for 12 h of daytime;} \]

\[ W_R = 5.28 \times 1.53 \times (30 - 15.5) \times 3600 \text{ ss} \times 12 \text{ h} = 117.14 \text{ J/s} \times 3600 \text{ ss} \times 12 \text{ h} = 5060.31 \text{ kJ for 12 h of nighttime.} \]

This assumes the yearly average of equal day and night length, and costs 1290.85 + 5060.31 = 6351.16 kJ for a statistically-average 24 h period.

Transfer of convection heat occurs between body surface and air temperature and its motion (Ingram and Mount, 1975). In calculation, the biggest difficulty takes place with determination of the size of \( h_C \) (convective heat transfer coefficient), due to this value fluctuating greatly as a result of its dependency on factors such as air temperature, the shape of body form, its size, etc. The most profound influence on \( h_C \) is had by wind speed, which sets the strength of the forced convective heat transfer. In the northern Kwa-Zulu Natal, wind speed averages at about 5 m.s\(^{-1}\) (Weather, 2017). At this wind speed, the value of \( h_C \) is about 660% greater than at the common indoor wind speed of 0.2 m.s\(^{-1}\) (Mitchell, 1974).

In an animal, convection occurs within the layer of exposed fur, which provides some insulation against the wind. However, at wind speed of 5m.s\(^{-1}\), about half of this insulation will be lost (McArthur, 1981).

Since no forced convective heat transfer coefficient has ever been determined for a wild felid species, the following is an approximation (determined for a sheep by Joyce et al. (1966)):

\[ h_C = 7.1v^{0.5}, \]

where \( v \) is the wind speed (in m.s\(^{-1}\)).

Substituting the wind speed of 0.2 m.s\(^{-1}\), we obtain a value of 3.18 for \( h_C \). In northern Kwa-Zulu Natal, considering the average wind speed frequent in the region, the value of \( h_C \) will be 660% greater: 3.18 \times 6.9 = 21.91. The precise quantity of convectional heat transfer can then be determined by the equation of Newton-Richman:

\[ W_C = h_C \cdot S \cdot (\delta_1 - \delta_2) \cdot t = 21.91 \times 1.53 \times (30 - 26.5) \times 3600 \text{ seconds} \times 6 \text{ hours} = 2534.29 \text{ kJ for 6 h of daytime activity.} \]

For nighttime, \( W_C \) becomes:

\[ 21.91 \times 1.53 \times (30 - 15.5) \times 3600 \text{ ss} \times 6 \text{ h} = 10499.18 \text{ kJ.} \]

If we assume that the African leopard is a predominantly nighttime hunter, performing most stalking and chasing
activities in the cooler temperatures, the animal will expend \( (2534.29*0.2)+(10499.18*0.8) = 8906.20 \text{ kJ} \). From existing literature, heat loss due to radiation and convection forms 73 to 88% of overall energetic heat losses (Ivanov, 1990); therefore it is best to assess the remaining heat-related energetic costs through relational means:

\[
W_E = \frac{(W_R + W_C)}{75}*25 = \frac{(6351.16 + 8906.20)}{75}*25 = 5085.79 \text{ kJ}
\]

From the aforementioned, the summative energetic costs attributed to heat transfer in general become:

\[
W_{\text{heat}} = W_R + W_C + W_E = 6351.16 + 8906.20 + 5085.79 = 20343.15 \text{ kJ}.
\]

Having accounted for both basal metabolism and heat transfer, it is important to consider the energy expenditure an African leopard is likely to incur while obtaining food. A leopard’s hunt consists of a number of stages (Stander et al., 1997): regular average-speed runs (10 km/h) to detect the presence of potential prey, a period of stalking and crouching, followed by a brief sprinting phase, during which speeds of up to 60 km/h (Nowak, 1999) are reached.

In case of the attempt being unsuccessful, the hunting process repeats itself over all. On average, every 1 in 4/5 hunts are successful (Bailey, 1993; Stander et al., 1997). When a prey item is killed, the leopard is likely to attempt relocating it to a competitor-free zone, dragging it at speeds of about 6 km/h (equating to about 1.7 m/s). Although there are known cases of \( P. \) pardus feeding on the same carcass for a few days, they often lose their kill after the initial feeding, frequently to hyenas (Creel et al., 2001).

Utilizing the aforementioned information, we can formulate a model of a standard African leopard hunt. The summative 10 km/h runs constitute a distance of about 8 km of daily movement. During this time, it initiates a maximum of 5 sprints as part of its hunting attempts, which together cover a distance of 200 m at 60 km/h speeds, an average of 40 m covered per sprint (Bothma, 1998).

In this model, energy losses associated with conversion from average to maximum speeds and back are not considered. The leopard’s ideal prey of 23 kg weight is dragged at speeds of 6 km/h for an average of 320 m (Smith, 1978). Therefore, average daily runs:

\[
W_1 = \frac{mv_1^2}{2}t_1 = 60 \text{ kg}*(2.8^2 \text{ m/s}^2)*2880 \text{ seconds} = 677.38 \text{ kJ}
\]

The maximum 5 hunting sprints per day (final one being successful):

\[
W_2 = \frac{mv_2^2}{2}t_2 = 60 \text{ kg}*(16.6^2 \text{ m/s}^2)*2*12.05 \text{ seconds} = 99.61 \text{ kJ}
\]

Relocation of the prey carcass to a safe feeding location:

\[
W_3 = (m_1+m_2)*v_2^2/2*t_3 = (60+23)*(1.7^2/2)*188 \text{ seconds} = 225.48 \text{ kJ}
\]

Summative daily expenditure of kinetic energy on movement activities:

\[
W_{\text{kinetic}} = W_1 + W_2 + W_3 = 677.38 + 99.61 + 225.48 = 1002.47 \text{ kJ}
\]

Summing the leopard’s overall daily energetic costs, we obtain:

\[
W = W_{\text{heat}} + W_{\text{kinetic}} + W_{\text{BM}} = 20343.15 + 1002.47 + 6048 = 27393.62 \text{ kJ}
\]

From the aforementioned, there is very little difference between the daily energy obtained (27625 kJ) and that used up on essential survival activities (27393.62 kJ). The small surplus of energy can either be used on particular activities which do not form part of daily routine (such as mating, territory defense, or unusually lengthy movement due associated with relocation), or can be retained for use the following day. Therefore for a statistically average African leopard, a sustained energetic balance is possible if, and only if, the biomass of prey is maintained at a level high enough to allow for successful food acquisition after at most the 5th hunting attempt. Let us consider the probability of such a situation. Marking every successful hunting sprint of a leopard with the variable \( p \), the probability of a successful hunt after \( n \) attempts can be presented as:

\[
R_{1,n} = 1 - q^n
\]

where \( q = 1 - p \), and is the chance of failure.

Using this equation, the values of \( p \) can be tested from 0.5 to 0.05, alongside varying \( n \) values, the results of which are presented in Table 1. As can be deduced from Table 1, under a success rate probability of \( p = 0.5 \), practically four or five hunting attempts are required before a kill is almost guaranteed to be made (chances of a kill being above 0.9). Under conditions of \( p = 0.3 \), the number of attempts required for the same effect almost doubles; while at \( p = 0.1 \) there need to be at least 20 hunts to allow for at least one success, a hunting requirement that cannot be adequately met on average, if using the energetic balance determined earlier.

**DISCUSSION**

The results indicate that under statistically average
leopard requirements in the African savanna, the probability of a successful kill needs to be at least $p = 0.5$ to ensure the persistence of the animal in the region (Table 1). This is due to the probability reflecting the necessary number of hunting attempts before a kill is guaranteed, which approximate four or five under $p = 0.5$ (a maximum of five hunting sprints being the basis on which the energetic calculations were made, which themselves presented a near-equilibrium between the energetic gains and losses). At lower probabilities, such as $p = 0.3$, the hunting effort required may lead to some levels of African leopard depopulation in the affected region, due to it being energetically unsustainable for the entire population to undertake almost eight hunts on average. However, it is the $p = 0.1$ scenario that is of most concern ecologically: a leopard’s energy reserves are insufficient to allow at least 20 hunts on average for one to be successful, therefore predicting an absolute removal of the leopard population from the region.

From such findings, it is now possible to determine the actual stocking rates needed to sustain a leopard population of a chosen size, by using the derived probabilities of hunting success and working with the specific energy content of meat from a prey species of interest. This may be highly beneficial to game reserve managers, and important in leopard conservation overall as it allows for the reduction in the number of farmer-leopard conflicts, by meeting the leopards’ energetic needs and in-turn reducing the need for leopards to leave the reserve in search of supplementary prey.

However, it is important to note that following this stocking strategy would be unlikely to eliminate all potential issues that leopards can face in a closed reserve space, with problems such as inbreeding depression and the spread of disease being common in confined felid populations (Kettles and Slotow, 2009), and not exclusively influenced by the prey number.

This paper reflects on the need to consider animal energetics when determining conservation action, showing that even a minor change in prey abundance can have potentially disastrous outcomes for predators relying on it. While there are other solutions to maintaining an adequate energy intake which predators often adopt in challenging circumstances – such as increasing the amount of food consumed per catch if prey is encountered less frequently; using the proportion of successful hunts is a convenient way to measure African leopard prey stocking rates. With habitat for leopards and their natural prey progressively declining solely to areas designated specifically for biodiversity preservation, determining successful hunts and adjusting reserve prey stocking rates in accordance will aid in maintaining stable leopard populations, and do much to alleviate farmer hostility where this is a problem, as mentioned above.

Combined with measures such as implementing alternative husbandry techniques for livestock keep and strengthening of policies guiding predator control on agricultural lands (Balme et al., 2009), leopard population reduction due to lack of prey can be effectively brought under control. With *P. pardus* research not consistently aligning with conservation priorities (Balme et al., 2013), there is a need for more applied studies addressing the factors controlling (and limiting) leopard occurrence, to effectively manage the distribution of *P. pardus* at large.

**CONFLICT OF INTEREST**

The author has not declared any conflict of interest.

**REFERENCES**

Bailey TN (1993). The African leopard: ecology and behaviour of a solitary felid. Columbia University Press, New York, USA.

Balme GA, Slotow R, Hunter LTB (2009). Impact of conservation interventions on the dynamics and persistence of a persecuted leopard (*Panthera pardus*) population. Biol. Conserv. 142:2681-2690.

Balme GA, Slotow R, Hunter LTB (2010). Edge effects and the impact of non-protected areas in carnivore conservation: leopards in the Phinda–Mkhuze Complex, South Africa. Anim. Conserv. 13:315-323.

Balme GA, Lindsey PA, Swanepoel LH, Hunter LTB (2013). Failure of research to address the rangewide conservation needs of large carnivores: leopards in South Africa as a case study. Conserv. Lett. 7:3-11.

Bellard C, Bertelsmeier C, Leadley P, Thuiller W, Courchamp F (2012). Impacts of climate change on the future of biodiversity. Ecol. Lett. 15:365-377.
Molecular view of trait

Accessed on: 31 March 2017

Bothma JP, Le Riche EAN (1986). Prey preference and hunting efficiency of the Kalahari Desert leopard In: Cats of the world: Biology, Conservation and Management, SD Miller, DD Everett (eds.), National Wildlife Federation, Washington DC, Pp. 389–414.

Bothma JDP (1998). Carnivore Ecology in Arid Lands. Springer-Verlag Berlin Heidelberg, pp. 7-42.

Burkey TV (1995). Extinction rates in archipelagos: implications for populations in fragmented habitats. Conserv. Biol. 9:527-541.

Chapman S, Balmé GA (2010). An estimate of leopard population density in a private reserve in KwaZulu-Natal, South Africa, using camera-traps and capture-recapture models. South Afr. J. Wildl. Res. 40:114-120.

Creel S, Spong G, Creel N (2001). Interspecific competition and population biology of extinction-prone carnivores. In Carnivore Conservation. Gittleman, J.L., Funk, S.M., Macdonald, D. and R.W. Wayne (eds.). Cambridge University Press, Cambridge, UK, pp. 35-61.

Deha K, Athreya V, Odden M, Linnell J (2012). Chemical immobilization of leopard Panthera pardus in the wild for collaring in Maharashtra, India. J. Bombay Nat. Hist. Soc. 109:153-157.

Fahrig L (1997). Relative effects of habitat loss and fragmentation on population extinction. J. Wildl. Manag. 61:603-610.

Fahrig L (2001). How much habitat is enough? Biol. Conserv. 100:65-74.

Fanger PO (1970). Thermal Comfort: Analysis and Applications in Environmental Engineering, Danish Technical Press, Copenhagen, Denmark.

Fretwell SD (1987). Food-chain dynamics: the central theory of ecology. Oikos 50:291-301.

Giam X, Bradshaw CJA, Tan HTW, and Sodhi NS (2010). Future habitat loss and the conservation of plant biodiversity. Biol. Conserv. 143:1594-1602.

Gonini L, Linnell JDC, May R, Panzacchi M, Boitani L, Odden M, Nilsen EB (2012). Habitat heterogeneity and mammalian predator-prey interactions. Mammal Rev. 42:55-77.

Gurevitch J, Padilla DK (2004). Are invasive species a major cause of extinctions? Trends Ecol. Evol. 19:470-474.

Hayward MW, Henschel P, O’Brien J, Hofmeyr M, Balme G, Kerley GHI (2006). Prey preferences of the leopard (Panthera pardus). J. Zool. 270:157-163.

Hoffman LC (2000). The yield and carcass chemical composition of impala (Aepyceros melampus), a southern African antelope species. J. Sci. Food Agric. 80:752-756.

Ingram DL, Mount LE (1975). Man and Animals in Hot Environments. Springer-Verlag, New York, USA, pp. 10-14.

Ivanov KP (1990). Fundamentals of Energy in the Body: Theoretical and Practical Aspects. T.I. General energy, heat exchange and thermoregulation (Russian). Hayyax (Nauka), Leningrad.

Jacobson AP, Gemberg P, Lemeris Jr JR, Schoonover RF, Anco C, Breitenmoser-Würsten C, Durant SM, Farhadinia MS, Henschel P, Kamler JF, Laguardia A, Rostro-Garcia S, Stein AB, Dollar L (2016). Leopard (Panthera pardus) status, distribution, and the research efforts across its range. Peer. J. e1974, DOI 10.7717/peerj.1974.

Joyce JP, Blaxter KL, Park C (1966). The effect of natural outdoor environments on the energy requirements of sheep. Res. Vet. Sci. 3:742-359.

Karanth KU, Nichols JD, Kumar NS, Link WA, Hines JE (2004). Tigers and their prey: predicting carnivore densities from prey abundance. PNAS 14:4854-4858.

Kettles R, Slотов R (2009). Management of free-ranging lions in an encroached park reserve. South Afr. J. Wildl. Res. 39:23-33.

Kissiu BM (2008). Livestock predation by lions, leopards, spotted hyenas, and their vulnerability to retaliatory killing in the Maasai steppe, Tanzania. Animal Conserv. 11:422-432.

Martin RB, de Meulenaer T (1988). Survey of the status of the leopard (Panthera pardus) in sub-Saharan Africa. CITES, Switzerland.

McArthur AJ (1981). Chapter 3: Thermal insulation and heat loss from animals. In: Environmental Aspects of Housing for Animal Production, J.A. Clark (ed.). Butterworths, London, UK, P 47.

Mitchell D (1974). Convective heat transfer from man and other animals. In: Heat Loss from Animals and Man: Assessment and Control, J.L. Monteith, L.E. Mount (eds.). Butterworths, London, UK, P 72.

Miththakala S, Seidensticker J, O’Brien SJ (1996). Phylogeographic subspecies recognition in leopards (Panthera pardus): molecular genetic variation. Conserv. Biol. 10:1115-1132.

Norton P (1990). How many leopards? A criticism of Martin and de Meulenaer’s population estimates for Africa. South Afr. J. Sci. 86:218.

Nowak RM (1999). Walker’s mammals of the world (6th edition). Johns Hopkins University Press, Baltimore, Maryland, USA.

Noël K, Jackson PR, Cornillon F, and Rodriguez J (2001). Carnivore Conservation and biodiversity monitoring and conservation action plan. International Union for Conservation of Nature and Natural Resources/ Species Survival Commission Cat Specialist Group, Gland, Switzerland.

Parker KL, Barboza PS, Gillingham MP (2009). Nutrition integrates environmental responses of ungulates. Funct. Ecol. 23:57-69.

Preisser EL, Bolnick DI, Benard MF (2005). Scarred to death? The effects of informatie extinction and consumption in predator-prey interactions. Ecology 86:501-509.

Ray JC, Hunter L, Ziegouris J (2005). Setting conservation and research priorities for larger African carnivores. WCS Working paper no. 24. Wildlife Conservation Society, New York, USA.

Rodewald AD, Kearns LJ, Shustack DP (2011). Anthropogenic resource subsidies decouple predator-prey relationships. Ecol. Appl. 21:936-943.

Schmidt-Nielsen K (1984). Scaling: why is animal size so important? Cambridge University Press, Cambridge, UK.

Schönfeldt H (1993). Nutritional content of venison. In: Forum: The venison industry, Research requirements and possibilities. Irene Industrial Research Council, Pretoria, South Africa. pp. 51-60.

Seabloom EW, Dobson AP, Stoms DM (2002). Extinction rates under nonrandom patterns of habitat loss. PNAS 99:11229-11234.

Shell D (2001). Conservation and biodiversity monitoring in the tropics: realities, priorities, and distractions. Conserv. Biol.15:1179-1182.

Simmersoff D (1984). Mass extinction and the destruction of moist tropical forests. Журнал общей биологии (Zhurnal Obshchei Biologii) 45:767-778.

Skead CJ (2007). Historical incidence of the larger land mammals in the broader eastern Cape (2nd edition). Centre for African Conservation Ecology, Nelson Mandela Metropolitan University, Port Elizabeth, South Africa.

Smith RM (1978). Movement patterns and feeding behaviour of the leopard in the Rhodes Matopos National Park, Rhodesia. Carnivore 1:58-69.

Stander PE, Haden PJ, Kaegece II, Ghau II (1997). The ecology of asociality in Namibian leopards. J. Zool. 242:343-364.

Stein AB, Athreya V, Gemberg P, Balme G, Henschel P, Karanth U, Miquelle D, Rostro-Garcia S, Kamler JF, Laguardia A, Khorozyan I, Ghodousi A (2016). Panthera pardus. The IUCN Red List of Threatened Species 2016:e.T159554A102421779.

Stephens DW (2008). Decision ecology: foraging and the ecology of animal decision making. Cogn. Affect. and Behav. Neurosci. 8:475-484.

Swartpoel LH, Lindsey P, Somers MJ, van Hoven W, Dalerum F (2013). Extent and fragmentation of suitable leopard habitat in South Africa. Anim. Conserv. 16:41-50.

Switzer PV (1993). Site fidelity in predictable and unpredictable habitats. Evol. Ecol. 7:533-555.

Tilman D, May RM, Lehman CL, Nowak MA (1994). Habitat destruction and the extinction debt. Nature 371:65-66.

Werner, E., Peacor S. (2003). A review of trait-mediated indirect interactions in ecological communities. Ecology 84:1083-1100.

Weather SA (2017). Forecast for The Big 5 False Bay represented by Hluhluwe. http://www.weathersa.co.za/city-pages/
Whitfield DP (2003). Predation by Eurasian sparrowhawks produces density dependent mortality of wintering redshanks. J. Anim. Ecol. 72:27-35.

Wikelski M, Cooke SJ (2006). Conservation physiology. Trends Ecol. Evol. 21:38-46.