INTRODUCTION

Green areas in increasingly urbanized cities are becoming an essential element in urban planning. Nowadays, revitalisation and protection of green areas, including historic parks, involves the need to respond to changing spatial conditions [Adamiec and Trzaskowska 2012, Fortuna-An- toszkiewicz 2012]. The negative changes in park soils due to structural degradation and nutrient leaching should be prevented. The comprehensive study of soil in urban areas is one of the main research directions in environmental geochemistry. According to Konstantinova et al. [2019], the geochemistry of urban soils depends on a combination of natural and anthropogenic factors specific to each urban environment. Scharenbroch et al. [2005] emphasised that as the initial disturbance to the soil environment increases, the impact of urbanisation is reduced by the processes that improve the physical, biological and chemical properties of the soil, and of the soil-forming factors, time plays he most important role in these transformations.

Assessing the quality of soils is not easy because of their complexity as well as the variability of their physical, chemical and biological properties and conditions. The sorption properties of soils play an important role as a factor regulating the leaching of nutrients from the soil, determining the efficiency of fertilisation and regulating plant nutrition [Meller et al., 2013]. Greinert [2009] believes that improving the sorption capacity of urban soils reduces the bioavailability of accumulated pollutants and is a prerequisite for the maintenance of urban green areas. Siuta [1995] reported that soil resistance to degradation is dependent on the sorption capacity of the soil and the content of alkaline cations in the soil. An important aspect in soil quality research is the relationship between the content of contaminants retained by the solid phase of the soil and the concentration of the contaminant in the soil solution.

The content of pollutants present in the soil solution is of great ecological importance due to the possibility of certain components being absorbed by plant roots. Contaminants such as e.g. heavy metals and PAHs, due to their properties,
pose a high potential risk to human health and may adversely affect the biotic elements of the soil environment.

The studies related to the assessment of the chemical properties of urban park soils are essential for taking the protective measures against the degradation resulting from inappropriate management.

The aim of the study was to evaluate the sorption properties of the anthrosols of the Park Ludowy before restoration, showing the current state of the soil and paying attention to the transformation of the soil cover, related to the park use and restoration works.

MATERIAL AND METHODS

Park Ludowy was established in the 1950s, it was one of the largest (31 hectares) and most important green areas in Lublin. It was established on wetlands on the right bank of the Bystrzyca River. The park was designed by an outstanding architect, Władysław Niemirski, and served primarily aesthetic, recreational and leisure functions.

In the 1980s, the significance of the park began to decline and it was subject to destruction and devastation. In addition, the lack of land reclamation and the dumping of soil from the settlers of the Lublin Sugar Factory into the park led to unfavourable changes in the root zone of the trees, causing their death. About 600 trees were cut down at that time. Attempts have been made to restore the beauty of the park, in the years 2000–2001 drainage works were carried out, which resulted in an improvement in the water properties of the soil, and in 2002, in order to make this not very representative place more attractive, the first facility of the Lublin International Fair was built in the park. Currently, the park is in the vicinity of other recreational facilities such as Arena Lublin, the Olympic swimming pool Aqua Lublin, the athletics stadium at al. Zygmunta and the Lublin Trade Fair (Fig. 1). In 2015, a project to revitalise the Park Ludowy was developed on the basis of a community concept and in collaboration with its authors. Even at the stage of design works, the idea was presented to residents and submitted for public consultation. The revitalisation began in 2019 and included comprehensive

Figure 1. Map of the Park Ludowy in Lublin [https://www.google.pl/maps/@51.2353739,22.5598855,16.33z]
landscaping of the greenery, as well as creation of didactic gardens and educational nature trails. The boundaries of the Park Ludowy are defined by al. Józefa Piłsudskiego to the east, ul. Lubelskiego Lipca 80 to the south, ul. Stadionowa to the west and the Bystrzyca River to the north.

The main objective of the Park Ludowy revitalization is to protect biodiversity, to restore natural resources, to organize build recreational and educational infrastructure and to make the park space available to residents while preserving and promoting the natural heritage. The activities undertaken within the framework of the investment are primarily aimed at effective protection of the ecosystems existing in the Park Ludowy, stopping the degradation of soil and urban greenery and contributing to shaping the pro-ecological attitudes in society. The renewal of the greenery in the park is intended not only to prevent the division of green areas, but also to create links to existing natural zones in the valley of the Bystrzyca River. [Resolution of the City Council 2017].

The anthropogenic soils of the Park Ludowy – one of the most important parks in the centre of Lublin – were studied. According to the IUSS Working Group WRB [2015], these soils are classified as Terric Anthrosols, and according to the Systematics of Soils of Poland [PTGleb. 2019], as anthropogenic soils order, culturoterrestrial soil type, rigosols subtype. Three of the samples analysed met the organic material criterion [IUSS soil type, rigosols subtype. Three of the samples as anthropogenic soils order, culturoterrestrial Systematics of Soils of Poland [PTGleb. 2019], were created as a result of mixing peat and silt material with materials of anthropogenic origin. In the studied soils of the Park Ludowy, the introduction of float soil and loess had a significant effect on the sorption properties, which varied the basic parameters of the soil material, both between and within pedons [Jaroszuk-Sierocińska and Słowińska-Jurkiewicz 2018]. The authors

The following determinations have been made in accordance with the methodology adopted in soil science studies: the organic carbon content was determined using an organic carbon analyser – (TOC) Shimadzu Corporation using the SSM-5000A solid sample combustion unit, the soil reaction (pH in 1 mol KCl-dm-3) was tested potentiometrically, the content of calcium carbonate (CaCO₃) was measured using the Scheibler’s method, hydrolytic acidity (Hh) was measured using the Kappen’s method, the sum of base cations (S) in non-carbonate samples using the Kappen’s method, and in carbonate samples using the Pallmann’s method. Cation exchange capacity(T) and the degree of saturation of the sorption complex with basic cations were calculated (Vₐ). The results of the analysis of sorption properties of the Park Ludowy soils were statistically analysed. An analysis of variance was performed for double orthogonal classification (pedon × layer). The significance of differences between the means was determined using Tukey’s test (α = 0.05).

RESULTS AND DISCUSSION

The present soil cover of the Park Ludowy was created as a result of mixing peat and silt material with materials of anthropogenic origin. In the studied soils of the Park Ludowy, the introduction of float soil and loess had a significant effect on the sorption properties, which varied the basic parameters of the soil material, both between and within pedons [Jaroszuk-Sierocińska and Słowińska-Jurkiewicz 2018]. The authors

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Table 1. Basic chemical properties of antropogenic soils in Park Ludowy

| Pedon | Layer (cm) | Organic carbon g·100⁻¹·g⁻¹ | CaCO₃ | Reaction (pH KCl) |
|-------|------------|-----------------------------|-------|-------------------|
| I     | 0–25       | 15.00                       | 20.76 | 6.9               |
|       | 25–50      | 8.04                        | 18.84 | 6.9               |
|       | 50–75      | 28.76                       | 0.00  | 6.3               |
| II    | 0–25       | 11.16                       | 8.34  | 6.8               |
|       | 25–50      | 2.46                        | 8.28  | 6.8               |
|       | 50–75      | 14.64                       | 5.80  | 6.9               |
| III   | 0–25       | 8.64                        | 2.98  | 6.9               |
|       | 25–50      | 3.96                        | 2.51  | 7.1               |
|       | 50–75      | 20.82                       | 3.59  | 6.6               |
| IV    | 0–25       | 2.94                        | 0.00  | 6.7               |
|       | 25–50      | 3.72                        | 4.62  | 7.2               |
|       | 50–75      | 39.12                       | 0.00  | 6.1               |
| V     | 0–25       | 2.42                        | 2.54  | 7.2               |
|       | 25–50      | 1.01                        | 7.20  | 7.2               |
|       | 50–75      | 9.54                        | 13.64 | 7.1               |
[2018] showed that the addition of earthy materials reduced the organic carbon content of the park soils, but this was compensated by an increased proportion of fine mineral particles. Multiple anthropogenic transformations have resulted in transformations of both physical and chemical properties of these soils. The natural use of waste, such as floating soil, consists in using it to shape the biologically active surface of the earth, primarily to maintain and increase soil fertility [Siuta 2001]. Resztel et al. [1998] showed that floatation soil can be used to improve the richness of medium and good formations located in the vicinity of sugar factories. The results of their experiments show that the application of soil to peat-moorsh soils increased crop yields. The addition of the floating soil significantly improved the physical properties and pH of the soil, as well as its abundance. It follows that the agricultural use of floatation soil is an appropriate use for this waste. As the soil becomes richer in highly sorptive materials, its sorption properties are changed. Bielińska and Moczek [2010] stress that soil sorption in urban park areas is improved by bringing in organic matter from waste such as municipal waste composts or sewage sludge. Given the important role of soils in the surrounding space, Jakubus [2015] indicated changes in soil properties expressed by an increase in pH values and a favourable balance of organic matter. These phenomena should be seen as positive processes of importance for both agricultural production and environmental protection. In the studied soils, the pH of the majority of samples was neutral, except for organic and non-carbonate material from the 50–75 cm layer of pedons I and IV, where it was weakly acidic (Table 1).

The hydrolytic acidity in the park soils ranged from 1.20 cmol(+)-kg⁻¹ to 3.90 cmol(+)-kg⁻¹ (Table 2). Its highest value was found in the superficial layer (0–25 cm) of pedon III, while the lowest value was found in the middle layer (25–50 cm) of pedon V. In most pedons, the values of the analysed trait were similar, in the surface layers they were higher than in the middle layers, and then they increased again in the layers at the depth of 50–75 cm. Only in pedon II there was a decrease in hydrolytic acidity with depth. The differences analysed were statistically significant. The results of hydrolytic acidity show the changes occurring in the soils, a similar relationship as noted in the pH of the soils studied (Table 1). A significant increase in hydrolytic acidity was found in the soils in which the pH value decreased.

In anthropogenic soils, elevated values for characteristics such as total base cations, saturation of the sorption complex with bases and reaction illustrate the typical phenomenon of mixing construction waste and lime-containing materials into the soil.

The properties of urban soils are largely shaped by reclamation measures carried out before the establishment of green areas [Greinert 2009]. Before reclamation, the sum of base cations in the soils of the Park Ludowy was quite high, ranging between 41.80 and 49.00 cmol(+)-kg⁻¹ (Table 2). Both these values occurred in pedon V in layers 0–25 cm and 50–75 cm. In the surface layers of most pedons, the values of the analysed trait were lower and increased with depth. Only in pedon III the sum of base cations in the middle layer was slightly higher (48.80 cmol(+)·kg⁻¹) than in the surface layer (48.40 cmol(+)·kg⁻¹). The smallest statistically significant difference at a significance level of α = 0.05 between the layers was 0.329 cmol(+)·kg⁻¹.

The higher capacity of the surface layers is typical of urban soils. In deeper layers in the soil profile, the sorption capacity often depends on the overlying organic and waste layers of porous materials [Greinert et al. 2013]. In a study by Bielińska and Mock [2010], the sorption properties of soils in urban parks were different, associated with asynchronous distribution of soil material during urban transformations in these areas. According to these authors [2010], the addition of organic compounds and mineral materials with significant sorption capacity to the soil environment under conditions of long-term urbanization is associated with the use of urban soils, moreover it involves the addition of both natural and anthropogenic substrates differing in composition, quantity, origin, and also in the way they are spatially displaced.

The cation exchange capacity of park soils ranged from 43.30% in pedon V to 52.35% in pedon II. In three pedons (I, IV and V) the values of the analysed property increased with depth, in the remaining two (II and III) the sorption capacity decreased with the depth of the profile and an increase of this property in the layer at the depth of 50–75 cm was recorded only in pedon III (Table 2).
A very important criterion for changes in the chemical properties of anthropogenic soils is the degree of saturation with alkaline cations. Końceka-Bentley et al. [1984] pointed out already in the 1980s that in chemically and geomechanically transformed soils the degree of base saturation reached 100% at alkaline pH and decreased to 90% further down in the profile. These authors showed that in urban soils with little transformation (large parks), the degree of base saturation was 60% in the humus level and increased up to 80% in the depth of the profile. In all the studied pedons of the Park Ludowy, the degree of saturation with alkaline cations was very high and varied on average from 93.902% in pedon II to 96.577% in pedon V and remained at the same level in all layers (from 94.087% to 95.862%). Within the layers in particular pedons, the highest degree of saturation with alkaline cations was observed in the second layer (25–50 cm) of pedon V – 97.590%, while the lowest was in the first layer (0–25 cm) of pedon III – 92.545%. Such values may occur at high intensity of anthropogenic factors and high degree of soil transformation and the abundance of alkaline cations in deeper layers may be higher than in the surface level. The analysis of variance showed statistical significance of the studied trait.

From the data discussed, it is evident that the anthropogenic factors that have influenced the soils of the Park Ludowy, such as the introduction of floatation soil and loess, have played a significant role in transforming soil processes, resulting in differences in their sorption properties, nutrient content, as well as in the structure of the soil profile.

**CONCLUSIONS**

The heterogeneous structure of the park showed that the different anthropogenic soils create new habitats and conditions favourable to the preservation of natural vegetation. Anthropogenic additives introduced into the soil in the form of float soil and loess had a beneficial effect on the sorption properties of soils in the root zone of plants. Hydrolytic acidity and total alkaline cations showed large or significant variation both with pedon and depth of layer. Cation exchange capacity and the degree of base cation saturation of the anthropogenic park soils remained very high and were characteristic

| Pedon | Layer (cm) | \(H_h\) | S | T | \(V_s\) |
|-------|------------|---------|---|---|--------|
|       |            | cmol(+)·kg\(^{-1}\) | % |   |        |
| I     | 0–25       | 3.03    | 47.22 | 50.25 | 93.962 |
|       | 25–50      | 2.25    | 48.28 | 50.53 | 95.542 |
|       | 50–75      | 3.22    | 48.60 | 51.82 | 93.780 |
| II    | 0–25       | 3.75    | 48.60 | 52.35 | 92.838 |
|       | 25–50      | 3.00    | 48.40 | 51.40 | 94.159 |
|       | 50–75      | 2.70    | 48.40 | 51.10 | 94.709 |
| III   | 0–25       | 3.90    | 48.40 | 52.30 | 92.545 |
|       | 25–50      | 1.95    | 48.80 | 50.75 | 96.158 |
|       | 50–75      | 3.00    | 48.42 | 51.42 | 94.167 |
| IV    | 0–25       | 2.55    | 44.20 | 46.75 | 94.545 |
|       | 25–50      | 2.10    | 48.60 | 50.70 | 95.859 |
|       | 50–75      | 3.00    | 48.80 | 51.80 | 94.199 |
| V     | 0–25       | 1.50    | 41.80 | 43.30 | 96.546 |
|       | 25–50      | 1.20    | 48.60 | 49.80 | 97.590 |
|       | 50–75      | 2.25    | 49.00 | 51.25 | 95.595 |

| Pedon | Layer (cm) | \(H_h\) | S | T | \(V_s\) |
|-------|------------|---------|---|---|--------|
| I     | 0–25       | 2.837   | 48.031 | 50.868 | 94.428 |
|       | 25–50      | 3.152   | 48.467 | 51.619 | 93.902 |
|       | 50–75      | 2.950   | 48.541 | 51.491 | 94.289 |
| II    | 0–25       | 2.551   | 47.201 | 49.752 | 94.868 |
|       | 25–50      | 1.650   | 46.433 | 48.083 | 96.577 |
|       | 50–75      | 2.946   | 46.046 | 48.992 | 94.087 |
| III   | 0–25       | 2.101   | 48.536 | 50.637 | 95.862 |
|       | 25–50      | 2.837   | 48.822 | 51.459 | 94.490 |
| IV    | 0–25       | 1.50    | 41.80 | 43.30 | 96.546 |
|       | 25–50      | 1.20    | 48.60 | 49.80 | 97.590 |
|       | 50–75      | 2.25    | 49.00 | 51.25 | 95.595 |

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|       | 50–75      | 2.25    | 49.00 | 51.25 | 95.595 |

| Pedon × layer | Means for pedons | \(H_h\) | S | T | \(V_s\) |
|---------------|-------------------|---------|---|---|--------|
| Pedons        | 0.258             | 0.499   | 0.499 | 0.611 |
| Layers        | 0.170             | 0.329   | 0.329 | 0.403 |

**Explanations:** \(H_h\) – hydrolytic acidity, S – sum of exchangeable cations, T – total sorption capacity, \(V_s\) – degree of basic cations saturation
of highly chemically transformed soils. After the revitalization of the park and before any new planting, special attention should be paid to the chemical properties of the soil cover and especially to its very strong sorption capacity.

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