Dispersion of Natural Airborne TiO$_2$ Fibres in Excavation Activity as a Potential Environmental and Human Health Risk

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1. Introduction

Titanium is the ninth most abundant element and constitutes approximately 0.7% of the Earth crust. The main titanium minerals are rutile (TiO$_2$), ilmenite (FeTiO$_3$) and titanite (CaTiSiO$_5$) [1].

Titanium dioxide (TiO$_2$) occurs as three polymorphs, rutile and anatase, which crystallize in the tetragonal system, and brookite, which crystallizes in the orthorhombic system. Rutile is an important accessory mineral present in felsic and mafic rocks, from a low to a high metamorphic degree, as well as a detrital mineral in sedimentary rocks. Typically, rutile may incorporate different elements such as Al, V, Cr, Fe, Zr, Nb, Sn, Sb, Hf, Ta, W and U [2].

Titanium dioxide is mined worldwide in large quantities for various applications such as paints, varnishes, lacquer, paper, plastic, ceramics, rubber, and printing ink. TiO$_2$ is also used in welding rod coatings, floor coverings, catalysts, coated fabrics and textiles, cosmetics, food colourants, glassware, pharmaceuticals, roofing granules, rubber tire manufacturing, and production of electronic components and dental impressions [3–6].

Titanium dioxide is chemically inert and exhibits photocatalytic activity. Therefore, the demand for TiO$_2$ in the global market has increased exponentially, causing intensive
extraction and consequent mineral release in the environment due to mining and industrial processing with potential environmental and human health effects. Current estimates calculate that in 2016 approximately 6.1 million tonnes of TiO$_2$ were used and this will reach 7.8 million tonnes by 2022 [7]. Moreover, different anthropic activity such as tunnel excavation, material handling, placement of stones and boulders can cause dispersion of TiO$_2$ into the environment in varying respirable particle-size fractions (diameter < 10 µm), including fine (diameter between 0.1 and 3 µm) and ultrafine (diameter ≤ 0.1 µm) sizes [8].

The International Agency for Research on Cancer (AIRC) includes TiO$_2$ in Group 2B as possibly carcinogenic to humans suggesting that pathological effects correlate to particle size and shape [9]. Albeit the National Institute for Occupational Safety and Health (NIOSH) in 1998 acknowledges TiO$_2$ as an occupational carcinogen, it does not recommend an exposure limit [10]. However, recently the NIOSH suggested airborne exposure limits of 2.4 mg/m$^3$ for fine TiO$_2$ and 0.3 mg/m$^3$ for ultrafine (including engineered nanoscale) TiO$_2$ as time-weighted average concentrations for up to 10 h/day during a 40-hour workweek [11].

Although different acts provide a quantitative risk assessment to ensure a safe and healthful working environment, currently the precise number of workers allegedly exposed to TiO$_2$ is inaccurate. Moreover, the main information on human health effects was principally obtained in animals by long-term, high-dose bioassay [12]. A few epidemiological studies have evaluated the carcinogenicity of TiO$_2$ exposed workers without reporting clear evidence of elevated risks of lung cancer mortality or morbidity from TiO$_2$ dust expositions [13–17]. On the other hand, these studies report a small number of deaths from a respiratory disease other than lung cancer and no report of pneumoconiosis associated with the inhalation of TiO$_2$.

A review of current toxicological data [18] has provided sufficient evidence for the carcinogenicity of titanium dioxide. Crystal dimensions may have a role in determining the level of toxicity [19]. Although the NIOSH establishes the highest acceptable concentration, different studies have demonstrated that the toxicity of TiO$_2$ particles primarily depends on the physical and mineralogical characteristics of the particles, such as size, specific surface coating and crystalline form [20–22]. Moreover, different authors have also investigated the effect of TiO$_2$ particles when inhaled or ingested [23]. In vivo studies reported that TiO$_2$ particles impact genotoxicity, pulmonary inflammation, oxidative stress by reactive oxygen species (ROS) and reactive nitrogen species (RNS) generation, lung tissue damage and cell proliferation that all concur in lung cancer development [24,25].

In recent years, particular attention has been paid to monitoring TiO$_2$ nanoparticle dispersion generated by industrial activities potentially harmful to human health. On the other hand, the goal of the present study was to quantify for the first time the release of natural TiO$_2$ particles in needle habit during excavation processes crossing meta-basalt and shale lithologies, determining the type, width, length, aspect ratio and concentration. In addition, this study highlights the importance of EMPs’ geometric relationships overlapping with the definition of fibres by the World Health Organization (WHO).

This activity can expose workers to natural respirable Elongated Mineral Particles (EMPs) during mining, geo-engineering and infrastructure construction. Moreover, it can increase the risk of exposure in the general population caused by the environmental dispersion of such EMPs.

2. Materials and Methods

In order to evaluate TiO$_2$ presence due to mining activity, samples from the Mt. Figogna Meta-basalts and Murta Shale formations were collected as representative lithologies crossed by the North-South oriented Terzo Valico tunnel across the Ligurian Alps, which is part of the Trans-European Transport Network (TEN-T) railway. The Mt. Figogna Meta-basalts originated in the Jurassic oceanic lithosphere and were overprinted under alpine pumpellyite-actinolite facies, developing the assemblage albite, Ca-amphibole, chlorite, epidote, pumpellyite and titanite, rutile and hematite as accessory mineral phases.
The Murta Shales is a low-grade metamorphic unit of the Ligurian Alps, made of dark grey shales with pervasive schistosity characterised by graphite-sericite alternating microlithons. Shales are locally rich in graphite or show decimetre-metre thick intercalations of micritic limestone and micaceous limestone. The mineral assemblage consists of mica, calcite, quartz, chlorite, while accessory minerals are magnetite or pyrite and rutile.

Five rock samples were gathered from the excavated rock front for each lithotype, and 10 airborne filter samples were obtained during excavation work. A total of ten rock samples and twenty airborne filters were analysed. The air sampling stations were located as close as possible to the source of dust during the excavation activity. Preparation and analyses were carried out at the Earth Environment and Life Sciences Department (DISTAV), University of Genoa, following methods prescribed for asbestos. In particular, the Italian legislative decree that regulates the protection of workers (L.D. 81/2008) refers to the method described in the Ministerial Decree (M.D.) 06/09/1994 (All. 2—Quantitative determination of airborne asbestos fibre concentrations in indoor environments), which is similar to the ISO Method 14996:2002 (ISO, 2002).

2.1. Rock and Airborne Sample Preparation

Airborne particulate was collected using a high flow sampling system depositing on polycarbonate membrane filters with a diameter of 25 mm and a pore size of 0.8 µm. Each sampling lasted for 8 h. Air was aspirated at a flow rate of 6 L/min, a total of 3000 L and was partitioned on 6 different filters due to a high concentration of environmental dust [26].

Differently representative aliquots (100 g) of rock samples were ground to a grain size of 100 µm with a Vibratory Disc Mill Retsch RS200 equipped with stainless steel grinding jars. Milling conditions were 25 s at 900 rpm and 15 s at 900 rpm for meta-basalt and shale samples, respectively. About 10 mg of the obtained powders were suspended in 100 mL of deionised water, and the obtained suspension was sonicated for 10 min. Five millilitres of the suspension was deposited on a polycarbonate membrane (47 mm diameter, 0.8 µm porosity) and filtered using a vacuum filtration system, obtaining about 0.5 mg of powders.

2.2. Scanning Electron Microscopy (SEM) analysis

All samples were analysed by SEM (Tescan Vega 3 XML), with a W source working at 20 kV of acceleration voltage. The elemental analysis was carried out by an Energy Dispersive X-ray spectrometer (EDS, Oxford Instruments, X-Max 20 SDD detector). A portion of each polycarbonate membrane was cut, mounted on an aluminium stub using a carbon conductive tab and coated with a thin gold layer by sputtering, using a Quorum Q150T ES.

For each sample, 54 random fields corresponding to a filter surface of 1 mm² were analysed by SEM-EDS at 2000× magnification. The number of TiO₂-EMPs was counted in each field, and the length and width were measured for each particle. Particle concentration in each sample was obtained applying the quantitative determination methods for airborne asbestos fibre using the Formula (1) reported in M.D. 06/09/1994, All. 2:

\[
C = N_f \frac{1}{a \cdot N_c \cdot A_f} \frac{1}{V}
\]

\(C = \text{particle concentration}\)
\(N_f = \text{total number of fibres counted}\)
\(a = \text{field area at 2000× (mm}^2\)\)
\(N_c = \text{total number of fields examined on the filter}\)
\(A_f = \text{effective collecting area of filter (mm}^2\)\)
\(V = \text{the volume of sampled air (L)}\)

The number of TiO₂-EMPs found was expressed as ff/L with confidence limits of 95%, assuming a Poisson distribution. Analytical sensitivity of the method was expressed, such as concentration, equivalent to observation of one fibre \((N_f = 1)\), and the \(V, A_f,\) and
The volume of each fibre was approximated to a cylinder, having diameter and height equal to fibre width and length, respectively. The weight of each EMP was calculated assuming an average density of the rutile (4.2 g/cm³) and correlated with the volume of particles. Moreover, the concentration expressed as mg/m³ was calculated assuming \( N_f \) as the total weight of the fibres found (mg) and the volume as the volume of sampled air expressed in m³.

2.3. Micro-Raman Spectroscopy

Micro-Raman analysis was performed by a Horiba Jobin-Yvon Explora-Plus spectrometer equipped with a charge-coupled device (CCD) detector, an Nd-YAG laser (532 nm), and Olympus BX 40 optical microscope with a 100X LWD objective and a grating of 2400 grooves/mm leading to a nominal spectral resolution of about 2 cm⁻¹. The spectrometer was calibrated to the silicon Raman peak at 520.5 cm⁻¹. Spectrum acquisition time was 20 s averaged between two accumulations in the spectral range between 50 and 1000 cm⁻¹. Analyses were carried on TiO₂ fibres sampled on a polycarbonate membrane filter.

3. Results

Meta-basalts and shales represent the principal lithotypes in the monitored area. In both lithotypes, the primary source of TiO₂ is represented by rutile, which exhibits an acicular and fibrous habit, as shown in Figure 1A,B.

![Figure 1](image-url)  
**Figure 1.** Detail of airborne rutile fibers in shale sample at 20,000× (A,B) and the representative elemental analysis (C). The results are shown as weight percentage (Wt%). The statistical error is displayed as sigma (σ), representing the overall confidence figure for the analysis, especially when an element is present at a low concentration. Scale bar in microphotographs; High Vacuum: 20 kV; Detector: Back Scattered Electrons and Energy Dispersive Spectroscopy.
The results obtained by μ-Raman (Figure 2) confirmed that the detected polymorph in both lithologies was rutile.

![Figure 2](image-url)

**Figure 2.** Raman spectrum of a needle (red) compared to a reference spectrum of rutile (black) from the free database Handbook of Raman spectra Lyon University (http://www.geologielyon.fr/Raman/spectrum accessed on 6 April 2021). Black dashed lines are eye-guides for the wavenumbers of the three Raman active modes in rutile (near 146, 447 and 612 cm\(^{-1}\)). Red dotted lines indicate two of the main Raman peaks visible from the surrounding polycarbonate (near 702 and 884 cm\(^{-1}\)).

Generally, the excavation procedure generates a large amount of dust from the rocks, which, disintegrating, release tiny particles suspended in the air. The extent of dust formation, as well as persistence in the environment, are relatable to the nature of the rocks and the size of the particles together with the environmental conditions such as humidity, temperature, circulating air and gas emissions inside the tunnel.

### 3.1. TiO\(_2\)-EMPs in Massive Rock Samples

The TiO\(_2\)-EMPs found in massive meta-basalts, despite being poor, had a length from 4.5 to 6.65 μm, averaging 5.68 μm. The average diameter is 0.38 μm with values ranging from 0.21 to 0.54 μm (Figure 3A). Noteworthy, the length to diameter ratio (L/D) of fibres in both lithotypes exceed the World Health Organization (WHO) established ratio of 3:1 [27].

In massive shale samples, TiO\(_2\)-EMPs showed a length of between 2.9 and 20 μm while the average value was 6.70 μm. All fine fibres detected had a diameter between 0.1 and 1.90 μm with an average diameter of 0.48 μm. The lengths of fibre distributions showed the following percentage: 36.84% for fibres from 2.5 to 5 μm, 33.33% between 5 to 7.5 μm, and 20.35% between 7.5 to 10 μm, respectively. The remaining 9.47% was distributed between lengths from 10 to 20 μm (Figure 3B).

Figure 4A,B shows examples of SEM analyses of rutile needle in the massive shale sample.
Figure 3. Size distribution of fibres in meta-basalt (A) shale (B) and massive samples.

Figure 4. Acicular rutile from massive shale sample. Overview of the reading field at 2000 × (A) and detail of rutile fiber at 20,000 × (B). Scale bar in microphotographs; High Vacuum: 20 kV; Detector: Back Scattered Electrons.

3.2. TiO$_2$-EMPs in Airborne Sample

3.2.1. Meta-basalts

SEM observation of the airborne samples obtained during the meta-basalt excavation front exhibited a fibre distribution whose length ranges between 3.24 and 14.06 µm, with an average length of 6.23 µm. About 69 TiO$_2$-EMPs were found, and of these 43.5% had a length of between 5 to 7.5 µm, while 36.2% showed a diameter lower than 0.4 µm. Moreover, 31.9% of the fibres reported a length from 2.5 to 5 µm and 18.8% from 7.5 to 10 µm, respectively. The remaining 5.8% varied between lengths from 10 to 20 µm. All fibres detected had a diameter between 0.2 and 1.6 µm with an average diameter equal to
0.48 µm and, therefore, were classifiable as fine fibres (Figure 5A). The results also indicated the presence of a variable aspect ratio where the prevalent classes were 10:1 (23.18%), 20:1 (46.38%) and 30:1 (13.04%). The meta-basalt fibres did not exceed ratios up to 40:1.

Airborne TiO$_2$ fibres concentrations in meta-basalt samples were 4.21 ff/L, ranging between 1.22 and 10.99 ff/L. The average concentration expressed as mg/m$^3$ was equal to 2.47·10$^{-5}$ mg/m$^3$ with values between 6.44·10$^{-5}$ and 2.85·10$^{-6}$ mg/m$^3$. TiO$_2$ fibre concentrations were expressed as both ff/L in the grey bar and mg/m$^3$ in a black diamond, as reported in Figure 6A.

3.2.2. Shales

SEM-EDS observations performed at 2000× magnification showed a significant number of TiO$_2$ fibres (392 fibres) from airborne shale samples with a length of between 2.9 and 20 µm with an average length of 7.04 µm. The 24% of fibres reported a length from 2.5 to 5 µm, the 42.3% between 5 to 7.5 µm and the 24.7% from 7.5 to 10 µm, respectively. The remaining 8% were distributed between lengths from 10 to 20 µm. All fibres detected had a diameter between 0.1 and 2.2 µm and, therefore, were classifiable as fine fibres, with an average diameter of 0.52 µm (Figure 5B).

A further investigation aimed at detecting the possible presence of ultrafine particles was performed under 5000× magnification (Figure 7), evidencing an additional class (15.81%) of particles with a length of between 1.6 and 2.5 µm. These fibres showed a diameter ranging between 0.1–0.5 µm. Aspect ratio shows prevalent classes 10:1 (24.23%), 20:1 (49.74%) and 30:1 (17.09%). In summary, 75.77% had an aspect ratio > 20:1 and 7.39% had an aspect ratio > 40:1. Noteworthy, a small percentage of fibres reached ratios of up to 70:1.
Figure 6. TiO\textsubscript{2} elongated particle concentrations reported as mg/m\textsuperscript{3} (black diamond) or ff/L (grey bar) ± fiducial limits (F 95%) in airborne samples of meta-basalt (A) and shale (B).

Figure 7. Overview of the reading field of airborne sampling from shales at 2000× (A) and 5000× (B). Scale bar in microphotographs; High Vacuum: 20 kV; Detector: Back Scattered Electrons.

The average airborne TiO\textsubscript{2} fibre concentration in shale samples was 23.94 ff/L, ranging between 9.16 and 34.81 ff/L. The average concentration expressed as mg/m\textsuperscript{3} was 2.07·10\textsuperscript{−4} mg/m\textsuperscript{3} with values between 3.45·10\textsuperscript{−4} and 1.24·10\textsuperscript{−4} mg/m\textsuperscript{3}. TiO\textsubscript{2} fibre concentration was expressed both as ff/L and mg/m\textsuperscript{3} (Figure 6B).

4. Discussion

Excavation activities at Fegino and Polcevera sites (Genoa, Italia) were addressed with reference to the “Mt. Figogna Meta-basalts” and “Murta Shales” formations. The air...
monitoring activity performed to detect asbestos began during the different excavation phases, evidencing a significantly associated occurrence of TiO$_2$-EMPs. SEM and µ-Raman analysis performed on membrane filters indicated that, in both meta-basalts and shales, the TiO$_2$ polymorph was rutile. TiO$_2$ crystals occurred with fibrous habit principally as single needles, in some cases as elbow-shaped twinned fibres. In shales, TiO$_2$ fibres are often included in mica or chlorite. All fibres have homogeneous size distribution despite higher concentrations in shales. Therefore, the features of TiO$_2$ fibres such as shape and size and the textural properties of the isolated or enclosed examples, together with the mechanical stress induced by the excavation process, may not influence their dimensions, concentrations, and dispersion in the environment. The different concentration values are likely linked to the primary (detrital) or secondary (metamorphic) rutile origin and its high mechanical resistance.

The majority of all measured rutile crystal sizes fell in the dimension interval established by WHO for fibre classifications following the geometric requisites criteria of fibres (length greater than 5 µm, diameter smaller than 3 µm, and length diameter ratio equal to or greater than 3:1) [27]. In particular, the percentage of EMPs that covered the geometrical aspect of fibre were 80% and 64.32% for meta-basalts and shale in massive rock samples, respectively. In the airborne samples, the percentage was 71% for meta-basalts and 85% for shale. Detailed counting at 5000× has reported the occurrence of elongated rutile particles with length <5 µm, assessing the presence of ultrafine TiO$_2$ particles.

These fibres, if dispersed in the environment, can be inhaled or ingested, playing a crucial role in the etiopathogenesis of different diseases. Although the toxicity of TiO$_2$ particles was low, lung tumours were demonstrated in animal study after two years of exposure to a high concentration of fine particles [12]. However, TiO$_2$ fibres can cross the gastrointestinal tract of rats, translocating in peripheral tissue and lymph node, respectively [28]. Rutile is therefore considered a chemically inert mineral that can have a carcinogenic effect attributed to lung overload.

The American Conference of Governmental Industrial Hygienists has assigned TiO$_2$ fine particles a threshold limit value of 10 mg/m$^3$ as a time-weighted average for a typical 8 h workday and a 40-h workweek [29]. Further indications from Occupational Safety & Health Administration (OSHA) indicate the permissible exposure limit for TiO$_2$ fine particle concentrations at 15 mg/m$^3$ [30]. Our case study found average concentrations of TiO$_2$ in airborne samples equal to 2.47 ± 10$^{-5}$ mg/m$^3$ for meta-basalt and 2.07 ± 10$^{-4}$ mg/m$^3$ for shale, values significantly lower than those reported above.

This monitoring carried out in the indoor working environment reported values below the limit indicated by the control agencies referred to above. However, it is important to underline the high abundance of TiO$_2$ fibres observed in the analysed samples. Although several studies have been carried out to determine the respiratory carcinogenicity of TiO$_2$, the spherical shape of the mineral has been principally studied. This does not allow correlation of the significance of the impact of shape to disease development. The shapes of TiO$_2$ could be divided into two main types, spherical and fibrous particles. It is plausible that TiO$_2$-EMPs could have similar behaviour to asbestos fibres [31]. The carcinogenic mechanism of the fibres is traced back to their capacity to translocate to the pleura inducing malignant mesothelioma. Generally, fibrous particles induced lung tumours through persistent inflammation due to incomplete phagocytosis and a release of ROS and RNS, such as different proinflammatory cytokines [32,33]. Bio-persistence and poor solubility are believed to be the most critical factors in this toxicity.

The TiO$_2$-EMPs observed in airborne samples from meta-basalts and shale generally follow the size of fibres defined by WHO. Due to their small diameter (<3 µm), these can be classified as fine respirable fibres, able to reach the alveolar environment, triggering inflammatory and recruitment of inflammatory cells. In fact, the cellular damage induced by fibrous particles may occur both through direct nucleus damage, due to the penetration of EMP breaking the structure of DNA and generation of ROS, caused by metal species incorporated in rutile, such as Fe, Cr, and V able to trigger a redox-active reaction [34].
The inclusion of TiO$_2$ in group 2B is based on rat studies and chronic inhalation exposure. In this regard, different debates are ongoing such as the representativeness of rat lungs in comparison to those of humans. Moreover, it is not clear if TiO$_2$–EMPs reactivity is due to the chemical or physical (i.e., particle) nature. In fact, the biological mechanism leading to carcinogenicity induced by inhalation and ingestion of particles is only partly understood [12,35]. Different studies have suggested that TiO$_2$ might induce or promote colon tumours via inflammation and ROS production [36,37].

Currently, available information does not allow conclusions on the relationship between specific physico-chemical characteristics of TiO$_2$ particles and the mechanisms triggering pathological events [38]. Therefore, future studies should characterize particles thoroughly and focus on specific characteristics such as crystal structure, size, surface coating, and reactivity of TiO$_2$ in relevant cell types to investigate whether all forms of TiO$_2$ should be considered to be of equal toxic potential.

Air monitoring activity of the asbestos carried out at the Polcevera and Fegino sites highlighted a number of rutile fibres deriving from the process of excavation. These findings indicated the possible contamination of the environment, mainly due to the presence of rutile in the host rocks and anthropic activities performed at these sites. The EMPs released in the environment represent a secondary source of contaminants that can be conveyed in aquifers by percolating rainwater.

Jani et al. [28] reported rutile particle (sized 0.5 µm) uptake from the rat gastrointestinal tract and subsequent translocation to systemic organs after oral administration. TiO$_2$ particles, especially nanoparticles, have been demonstrated to cross the cellular barrier, both in vivo and in vitro, although sometimes translocation was mild and did not finalize in any significant systemic effects. Inflammation and enhanced initiation or promotion of colorectal carcinogenesis were observed but not confirmed [39].

Moreover, non-inhaled rutile fibres, despite their high specific weight, showed aerodynamic characteristics and persistence, remaining in the atmosphere for long periods, dispersing over long distances and contaminating different sites [40,41].

5. Conclusions

The present study results show that rutile environmental dispersion is mainly due to anthropological activity. TiO$_2$ fibres in the airborne samples have already been highlighted in our monitoring activities aimed at detecting the presence of asbestos minerals from the Polcevera and Fegino sites. The different samplings have reported that concentrations of TiO$_2$ never exceed the limits established by different organizations for workers’ health protection. Though the AIRC has included the TiO$_2$ in group 2B, it is noteworthy that various epidemiological studies do not support this classification. In this regard, new environmental and human biomonitoring studies would be necessary to facilitate the long-term risk assessment caused by inhalation or ingestion of elongated particles. This information may be helpful to increase scientific knowledge and promote reduced TiO$_2$ human exposure. Biomonitoring studies addressing the risk caused by TiO$_2$ inhalation in different work activities such as mining are currently lacking.

Contrarily to asbestos which has a proven harmful effect on human health, specific legislation on TiO$_2$ monitoring is lacking. However, the data discussed in this paper demonstrated that most TiO$_2$ elongated particles, recognized as rutile, had the dimensional characteristic of fibres, as reported by WHO. In particular, a fraction of the fibres fall is fine, able to reach alveolar space and trigger an inflammation process that can lead to pneumoconiosis. Pneumoconiosis may result in several complications such as tuberculous superinfection, chronic bronchitis, emphysema, respiratory failure, chronic pulmonary hypertension, and tumours of the respiratory system, especially in smokers, enhanced by a significant synergic effect.

The importance of monitoring TiO$_2$ in both working environments and naturally hazardous areas, such as stream sediments and internal waters across Ti-rich geological for-
mations, associated with epidemiological studies, may represent a useful tool to determine the exposure risk of workers and the whole population.

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