Oxidation of Crystal Violet in Aqueous Solutions Using the Fenton Process

Shohreh Heshmati¹, Zahra Karamizand¹, Hadis Fattahi¹ and Hooshyar Hossini², *

¹Students Research Committee, Kermanshah University of Medical Sciences, Kermanshah, Iran
²Department of Environmental Health Engineering, Faculty of Health, Kermanshah University of Medical Sciences, Kermanshah, Iran

*Corresponding author: Department of Environmental Health Engineering, Faculty of Health, Kermanshah University of Medical Sciences, Kermanshah, Iran. Email: hoo.hosseini@gmail.com

Received 2019 May 19; Revised 2019 June 20; Accepted 2019 July 06.

Abstract

Advanced oxidation processes (AOPs) are always accompanied by producing highly-reactive hydroxyl radicals (OH) with great potential for the decolorization and mineralization of organic compounds. Today’s, dyestuff and pigments are considered a notorious concern for the pollution of aqueous environments. Proposing new methods to dispose of this problem is therefore essential. The present study was conducted to investigate the use of the Fenton process for removing crystal violet (CV) from aqueous solutions. The effectiveness of this method was influenced by several factors, including pH, initial concentration of the dye, different interfering ions and reaction duration. The residual concentration of the dye was determined using spectrophotometry at a maximum wavelength of about 586 nm. The data were analyzed and interpreted in Excel. The optimal pH was obtained as 3, and the optimal initial concentration and Fe/H₂O₂ ratio as 50 mg/L and 1:10 respectively. Given the effect of the contact duration, the minimum Fenton reaction duration was considered 30 minutes. In next step the effect of interference in the progression of the Fenton process were investigated for KCl, NaCl and NaHCO₃. Moreover, the efficiency of CV decolorization was 96% under optimal conditions. Fenton can be therefore considered an appropriate process for removing CV from colored wastewater in textile industries.

Keywords: Colored Wastewater, Fenton Process, CV

1. Background

Crystal violet (CV) is a water-soluble cationic dye that is extensively used in dyeing industries for applications such as bio-staining, medicine, disinfection, foods (1). Large volumes of effluents produced in industrial processes can enter the environment as a result of widely using cationic dyes such as CV (2). The toxicity of these dyes is normally higher than that of anionic dyes. CV is extensively used in industries and, can enter the environment in large quantities through sewage. Using appropriate systems to safely dispose of this substance therefore appears essential (3). Over the years, CV has been proposed to be eliminated using different methods, including oxidative degradation (4), electrochemical techniques (5), membrane separation (6) and biochemical absorption and degradation (7). Among different refining processes, AOPs are very important in terms of decolorization and mineralization of organic compounds through generating hydroxyl radicals. Fenton is a cost-effective, safe, fast, highly-oxidative, non-energy intensive, simple and easy to operate and maintain (8) AOP that produces no byproducts. The main advantages of the Fenton process include combining oxidation and coagulation, which results in producing less sludge compared to that produced in coagulation and flocculation (9). In addition to eliminating organic matters (10), Fenton can be used for removing certain toxic elements (8) and for the pre-treatment of biological processes (11). The Fenton reaction involves several sequential reaction steps, which are described as follows (12):

1. H₂O → ∗OH + H⁺ + e⁻
2. O₂ + 2H⁺ + 2e⁻ → H₂O₂
3. Fe²⁺ + H₂O₂ → Fe(OH)²⁺ + ∗OH → Fe³⁺ + OH⁻ + H₂O
4. Fe³⁺ + H₂O₂ → Fe-OOH²⁻ + H⁺
5. Fe-OOH²⁻ → Fe²⁺ + HO₂⁻
6. H₂O₂ → HO₂⁻ + H⁺ + e⁻

The drawbacks of the Fenton process include the risk of storing and transporting hydrogen peroxide, dealing with high amounts of chemicals and the need for neutralization before disposing of wastewater (13); nevertheless, many researchers have reported the successful application of the Fenton process for colored wastewater and organic materials; for instance, Chen et al. used this process to eliminate...
surfactants, and achieved a removal efficiency of over 75% (11). Perkowski et al. used the Fenton process to treat high-density detergent wastewater with a removal efficiency of 65% - 97% (14). Meric et al. reported the elimination of over 99% of the reactive dye (BLACK 5) at a concentration of 200 mg/L (15). Biglari et al. showed that the Fenton process could be used to treat water containing organic soluble carbon (16).

2. Objectives

The present article seeks to investigate the catalytic decomposition of CV in aqueous solutions using the Fenton process.

3. Methods

3.1. Materials

CV is a trinyl-phenyl methane dye with a molecular formula of C_{24}H_{28}N_{3}Cl and a molecular weight of 396.96 g/mol. Figure 1 shows the chemical structure of CV.

![Figure 1. Chemical structure of CV](image)

3.2. Setup

A laboratory scale and a batch mode were designed for the decolorization of CV. A Pyrex glass cylinder with an effective volume 500 ml was used as the chemical reactor. A dye stock was prepared with an approximate 100 mg/L concentration of CV by dissolving the dye aliquots into tap water. Ferrous ions and hydrogen peroxide were used at a 1:10 ratio both with a concentration of 1 mM. To adjust the sample pH, 1 M NaOH and 1 M H_{2}SO_{4} were used in a WTW pH meter. A 100-rpm magnetic stirrer was also used to ensure the homogeneity of the solution. The maximum wavelength of CV was determined as 586 nm using a UV-VIS spectrophotometer (Jenway 6305). Data analysis and interpretation was performed in Excel 2010 (version 14.0).

4. Results and Discussion

4.1. Fe^{2+}/H_{2}O_{2}

Pollutants can be generally decomposed in the presence of iron ions and hydrogen peroxide owing to the formation of hydroxyl radical species, should the molar ratio be correctly selected. The initially-tested ratios of Fe^{2+}/H_{2}O_{2} included 1:10, 1:5 and 1:1. The optimal ratio of Fe^{2+}/H_{2}O_{2} maximizes the number of oxidative radicals produced. In other words, suboptimal doses of hydrogen peroxide act as a radical scavenger, and ultimately reduce the oxidation efficiency. In addition, high doses of hydrogen peroxide are practically used for the pre-disposal treatment of wastewater and deal with potential biological processes. The optimal doses of hydrogen peroxide and iron ions should be therefore obtained from laboratory studies. According to the results (data not shown), the optimal ratio was found to be 1:10.

4.2. Effect of pH

pH is an effective parameter in many chemical reactions. The optimal pH was determined by investigating its effect on the Fenton process through evaluating pH 3 and 5 as the indicator of the acidic character, pH 7 of the neutral character and pH 9 of the alkaline character. This test was conducted for 30 minutes at a CV concentration of 20 mg/L, a stirring speed of 100 rpm and an Fe^{2+}/H_{2}O_{2} ratio of 1:10. Figure 2 shows the effect of different pH values on the efficiency of the Fenton process in eliminating CV, suggesting that pH 3 is the optimal with an elimination efficiency of 96.1%. The catalytic activity of iron species highly depends on the solution pH. The concentration of divalent iron species and therefore the rate of the Fenton reaction and the decolorization percentage are maximized at pH 3. Moreover, at pH values over 5, trivalent iron species precipitate as Fe(OH)_{3} molecules, reducing the amount of catalyst in the solution. The amount of available free iron ions and hydroxyl radicals are reduced at higher pH values due to the deposition of Iron ions as Fe(OH)_{3} molecules. In other words, hydrogen peroxide is spontaneously decomposed at higher pH values.

A review of literature suggests that, at a pH below 3, the complex species of [Fe(H_{2}O)_{6}]^{3+} are formed and their
reaction with hydrogen peroxide is much slower than that of other species, which lowers the efficiency. At relatively high concentrations of $H^+$, i.e. a strong acidic pH, hydrogen peroxide is dissolved and forms stable oxonium ions of $H_3O^+$ (17). Given that the reactivity of oxonium ions with iron ions is much lower than that of hydrogen peroxide, the efficacy of the Fenton process in decomposing organic compounds is reduced at a very low or high pH, and the optimal pH of the Fenton process is 3.

4.3. Initial Concentration of Dyestuff

To determine the effect of dye concentration, 10, 25, 50 and 100 mg/L of CV were prepared. Figure 3 shows the effect of the initial concentration of the dyestuff on the Fenton process, suggesting a decrease in the efficiency of removing CV with increasing its concentration from 50 mg/L to 100 mg/L. At lower concentrations, especially below 50 mg/L, decolorization was successful with a minimum removal percentage of 86% - 96%. The Fenton process was therefore found to be more efficient at lower concentrations of the dye. These results suggest that dilution can be helpful or even necessary before performing the Fenton process. This finding can be explained by the fact that the decolorization of dye molecules is limited by the number of radicals and the contact duration. In other words, in lower concentration of dye, better decolorization performance is happened.

4.4. Effect of Interfering Salts

The effect of interfering salts such as KCl, NaCl and NaHCO3 were evaluated at this stage. This experiment was performed at a pH of 3, a $Fe^{2+}/H_2O_2$ ratio of 1:10, an initial concentration of 50 mg/L and a 0.1 g/L concentration of the interference. Figure 4 shows the effect of different salts on the efficiency of the Fenton process in removing CV, suggesting the significant impact of the interference, which is potentially associated with the relatively low concentration of the interference, i.e. about 0.1 g/L. The Fenton process can be therefore successfully accomplished in the presence of low levels of interfering salts. Research suggests that NaHCO3 exerts a significant scavenging effect on hydroxyl radicals. The bicarbonate ions can take the interfering role in the following reaction (18).

4.5. Literature Review

Table 1 summarizes the studies in which different dyes and pigments were degraded with Fenton and Fenton-based processes, suggesting that Fenton is a strong process for the degradation/decolorization of a wide range of dyes. Decolorization was completed in some cases within a few hours of operation. Research also suggests a dye mineralization efficiency of over 80% (19, 20).
Table 1. Comparing the Efficiency of the Fenton Process for Different Dissolved Dyes

| Process                      | Dye              | Wastewater | Experimental Condition                  | Efficiency   | Reference |
|------------------------------|------------------|------------|----------------------------------------|--------------|-----------|
| Fenton type: Fe^{2+}H_{2}O_{2}, Fe^{3+}H_{2}O_{2}, photo-Fenton based: UV/Fe^{2+}H_{2}O_{2}, (UV)/Fe(2H_{2}O) | Acid Orange 7    | Synthetic | Optimal: pH 3, Fe^{3+}H_{2}O_{2} 15 - 50, duration < 10 min, dye concentration: 20 mg/L | > 94%        | (21)      |
| Electro-fenton               | Azobenzene       | Synthetic  | pH 2 - 3, 0.5 mM ferric iron, Current: 60 mA | mineralization yield 80% | (19)      |
| Fenton, Fenton-like          | CV               | Synthetic  | Optimal: pH 3, 30 min, dye concentration: 0.5 mM, Fe^{3+}H_{2}O_{2} 0.5:50, Fe^{2+}H_{2}O_{2} 150 | ~100%        | (22)      |
| Fenton (H_{2}O_{2}/Fe^{3+}) and photo-Fenton (H_{2}O_{2}/Fe^{2+}/UV) | Reactive Black 5 | Synthetic | Optimal: [H_{2}O_{2}]/[RBS] 4.9:1, [H_{2}O_{2}]/[Fe^{2+}] 9.63:1, pH 3 | 97.5% 98.1% | (23)      |
| Heterogeneous Fenton-like    | Orange II        | Synthetic  | Optimal: pH 3, Catalyst: 0.2 g, H_{2}O_{2} concentration: 6 mM and temperature: 10°C | Mineralization 90% | (20)      |
| Electro-Fenton               | Alizarin red     | Synthetic  | Na_{2}SO_{4} 0.05 M, pH 3, T 25°C, Fe^{2+} 1 mM, Applied current: 200 mA, 120 min | ~100%        | (24)      |
| Fenton                       | CV               | Synthetic  | Optimal: pH 3, Duration: 30 min, Fe^{3+}H_{2}O_{2} 1:50, Dye concentration: 50 mg/L | > 96%  | Present study |   |

5. Conclusions

The present study determined the efficiency of the Fenton process in the decolorization of CV in an aqueous matrix. The optimal conditions included pH 3, an initial concentration of 50 mg/L and an Fe/H_{2}O_{2} ratio of 1:10. The obtained results suggest that the Fenton process can be used for the decolorization of colored wastewater.

Acknowledgments

The authors gratefully acknowledge the Research Council of Kermanshah University of Medical Sciences (grant number: 95492) for the financial support.

Footnotes

Authors’ Contribution: 25% for each authors.

Conflict of Interests: It is not declared by the authors.

Ethical Considerations: Ethical issues (including plagiarism, informed consent, misconduct, data fabrication and/or falsification, double publication and/or submission, redundancy, etc) have been completely observed by the authors.

Funding/Support: Research Council of Kermanshah University of Medical Sciences (grant number: 95492).

References

1. Idaka E, Ogawa T, Yatone C, Horitsu H. Behavior of activated sludge with dyes. Bull Environ Contam Toxicol. 1985;35(6):729-34. doi: 10.1007/bf01636580. [PubMed: 4074939].

2. Bagheri Marandi G, Baharlouli M. [Synthesis of hydrogel nannocomposites of acrylamide-taconic acid using laponite and study of crystal violet dye adsorption]. Iran J Polym Sci Technol. 2012; 24:505-14. Persian.

3. Jones J, Falkingham J0 3rd. Decolorization of malachite green and crystal violet by waterborne pathogenic mycobacteria. Antimicrob Agents Chemother. 2003;47(7):2323-6. doi: 10.1128/aac.47.7.2323-2326.2003. [PubMed: 12824499]. [PubMed Central: PMC306181].

4. Gozmen B, Kayan B, Girit AM, Hesenov A. Oxidative degradations of reactive blue 4 dye by different advanced oxidation methods. J Hazard Mater. 2009;168(1):129-36. doi: 10.1016/j.jhazmat.2009.02.011. [PubMed: 19269087].

5. Khansorthong S, Hunsom M. Remediation of wastewater from pulp and paper mill industry by the electrochemical technique. Chem Eng J. 2009;145(1-3):228-34. doi: 10.1016/j.cej.2009.02.038.

6. Purkait MK, Dasgupta S, De S. Micellar enhanced ultrafiltration of azo dye Direct Red 5B using Blumea malcolmii Hook. Bioreour Technol. 2009;100(18):4010-4. doi: 10.1016/j.biortech.2009.01.049. [PubMed: 19410449].

7. Kagalkar AN, Jagtap UB, Jadav JP, Bapat VA, Govindwar SP. Biotechnological strategies for phytoremediation of the sulfonated azo dye Direct Red 5B using Blumea malcolmii Hook. Bioreour Technol. 2009;100(18):4010-4. doi: 10.1016/j.biortech.2009.01.049. [PubMed: 19410449].

8. Lippczynska-Kochany E, Kochany J. Effect of humic substances on the Fenton treatment of wastewater at acidic and neutral pH. Chemosphere. 2008;71(5):745-50. doi: 10.1016/j.chemosphere.2008.06.028. [PubMed: 1857846].

9. Merc S, Selcuk H, Belgiorno V. Acute toxicity removal in textile finishing wastewater by Fenton’s oxidation, ozone and coagulation-flocculation processes. Water Res. 2005;39(6):1447-53. doi: 10.1016/j.watres.2004.12.021. [PubMed: 15766969].

10. Kavitha V, Palanivelu K. The role of ferrous ion in Fenton and photo-Fenton processes for the degradation of phenol. Chemosphere. 2004;55(9):1325-31. doi: 10.1016/j.chemosphere.2003.12.022. [PubMed: 15087654].

11. Chen S, Sun D, Chung JS. Treatment of pesticide wastewater by moving-bed biofilm reactor combined with Fenton-cosolagation pretreatment. J Hazard Mater. 2007;144(1-2):577-84. doi: 10.1016/j.jhazmat.2006.10.075. [PubMed: 17141410].

Int J Health Life Sci. 2019; 5(2):e94090.
12. Anotai J, Lu MC, Chewpreecha P. Kinetics of aniline degradation by Fenton and electro-Fenton processes. *Water Res.* 2006;40(9):1841–7. doi: 10.1016/j.watres.2006.02.013. [PubMed: 16624370].

13. Wun Jern NG. *Industrial wastewater treatment*. 2006.

14. Perkowski J, Jóźwiak W, Kos L, Stajszczyk P. Application of Fenton’s reagent in detergent separation in highly concentrated water solutions. *Fibres Text Eastern Eur.* 2006;59(5):14–9.

15. Meric S, Kaptan D, OZmez T. Color and COD removal from wastewater containing Reactive Black 5 using Fenton’s oxidation process. *Chemosphere*. 2004;54(3):435–41. doi: 10.1016/j.chemosphere.2003.08.010. [PubMed: 14575758].

16. Biglari H, Jonidi Jafari A, Kord Mostafapour F, Bazrafshan E. [Removal of dissolved organic carbon from aqueous solution by fenton oxidation process]. *J Birjand Univ Med Sci*. 2012;19(1):70–80. Persian.

17. Aramyan SM. Advances in fenton and fenton based oxidation processes for industrial effluent contaminants control-a review. *Int J Environ Sci Nat Resour*. 2017;2(4):1–18. doi: 10.19080/ijesnr.2017.02.555594.

18. Armstrong WA. Relative rate constants for reactions of hydroxyl radicals from the reaction of Fe(II) or Ti(III) with H2O2. *Can J Chem*. 1969;47(20):3737–44. doi: 10.1139/v69-623.

19. Guivarch E, Trevin S, Lahitte C, Oturan MA. Degradation of azo dyes in water by Electro-Fenton process. *Environ Chem Lett*. 2003;1(1):38–44. doi: 10.1007/s10311-002-0007-0.

20. Ramirez H, Maldonado-Hódar FJ, Pérez-Cadenas AF, Moreno-Castilla C, Costa CA, Madeira LM. Azo-dye orange II degradation by heterogeneous Fenton-like reaction using carbon-Fe catalysts. *Appl Catal B Environ.* 2007;73(3-4):312–21. doi: 10.1016/j.apcatb.2007.05.003.

21. Kusic H, Koprivanac N, Srsan L. Azo dye degradation using Fenton type processes assisted by UV irradiation: A kinetic study. *J Photochem Photobiol A Chem*. 2006;181(2-3):195–202. doi: 10.1016/j.jphotochem.2005.11.024.

22. Fan HJ, Huang ST, Chung WH, Jen JL, Lin WY, Chen CC. Degradation pathways of crystal violet by Fenton and Fenton-like systems: Condition optimization and intermediate separation and identification. *J Hazard Mater*. 2009;171(1-3):2032–44. doi: 10.1016/j.jhazmat.2009.06.017. [PubMed: 19604632].

23. Lucías M, Peres J. Decolorization of the azo dye reactive black 5 by Fenton and photo-Fenton oxidation. *Dyes Pigm.* 2006;71(3):236–44. doi: 10.1016/j.dyepig.2005.07.007.

24. Panizza M, Cerisola G. Electro-Fenton degradation of synthetic dyes. *Water Res*. 2009;43(2):339–44. doi: 10.1016/j.watres.2008.10.028. [PubMed: 18996554].