Environmental concentrations of metformin exposure affect aggressive behavior in the Siamese fighting fish, *Betta splendens*

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Abstract

Metformin, the medicine most commonly prescribed for treatment of Type II diabetes, is among the most abundant pharmaceuticals being introduced into the environment. Pharmaceuticals are increasingly found in wastewater and surface waters around the world, often due to incomplete metabolism in humans and subsequent excretion in human waste. Risk analyses and exposure studies have raised concerns about potential negative impacts of pharmaceuticals at current environmental levels. Results of the present study indicate that metformin at concentrations in the range of what has been documented in freshwater systems and waste-water effluent (40 μg/L) affects aggressive behavior in adult male *Betta splendens*. Subjects exhibited less aggression toward a male dummy stimulus after four weeks exposure to metformin-treated water when compared to behavior measured immediately prior to their exposure, and in comparison to a separate cohort of un-exposed control fish. This effect persisted after 20 weeks exposure as well. Subjects exposed to metformin at a concentration twice that currently observed in nature (80 μg/L) exhibited an even more substantial reduction in aggressive behaviors compared to controls and pre-exposure measurements than those observed in the low-dose treatment group. Such changes in behavior have the potential to affect male fitness and possibly impact the health of natural populations of aquatic organisms exposed to the drug.

Introduction

Anti-depression medications, birth control pills, among other pharmaceuticals are found in waste- and surface waters around the world in increasing concentrations due to incomplete metabolism in humans and excretion in waste [1–3]. These chemicals become part of wastewater treatment plant (WWTP) effluent released into local surface waters including streams, rivers, coastal estuaries [4,5]. A growing body of research has raised concerns about potential negative impacts of such pharmaceuticals at current environmental levels [1,6–8].

Metformin, the medicine most commonly prescribed by doctors worldwide for the treatment of Type II diabetes, is among the most abundant pharmaceuticals being introduced into the environment. The drug is found in WWTP effluent at concentrations of 1 μg/L to 47 μg/L.
and in surface waters at concentrations from 0.06 μg/L to 3 μg/L [1,3,9,10]. Levels of metformin in WWTP effluent and surface waters will likely continue to rise given its use as a treatment option for a broadening variety of health problems from cancers [11] to polycystic ovarian syndrome [12] and the staggering rate of increase in diabetes diagnoses in recent decades [3].

The effects of pharmaceuticals on wildlife have received increased attention in the literature in recent years, with particular emphasis on those that affect endocrine function. However, metformin’s impact on aquatic life is poorly understood, having been addressed in a limited number of studies exploring its metabolic effects in the contexts of aquaculture [13–16] and drug screening [17]. A recent study of fathead minnows (Pimephales promelas) exposed to the drug revealed that metformin treatment induced molecular genetic changes indicative of endocrine disruption. Males exposed to the drug for a period of four weeks at a level similar to the average found in WWTP effluent (40 μg/L) experienced significant up-regulation of mRNA encoding the egg-protein vitellogen (VTG), a well-established indication of endocrine disruption [8]. The authors suggest these changes offer the potential for greater hormone-, health- and behavior- disrupting impacts from extended exposure, as occurs in nature. Studies of the effects of metformin at higher doses, over longer periods of exposure, and in other species [18] are needed in order to better assess impacts of the drug in nature [8].

To date, no one has addressed the effects of metformin exposure on the behavior of aquatic life. Many animals may experience behavioral changes as a consequence of chemical exposure even in the absence of obvious morphological effects [19]. An assessment of behavior at various points in time during the course of an extended period of chemical exposure may provide important information on when and for how long these chemicals impact the health of aquatic organisms [19].

The Siamese fighting fish (Betta splendens) has long been a popular model for studying reproductive and aggressive behavior [20] and has become a model organism for studies addressing the physiological and behavioral effects of chemical exposure [19,21]. B. splendens are found in freshwater ponds of Southeast Asia and are available in pet stores worldwide. Both wild and captive-bred males exhibit discrete and stereotyped aggressive behaviors in defending their territories against intruding males [22]. The aim of the present study was to assess whether chronic exposure of adult male B. splendens to waterborne levels of metformin comparable to those found in WWTP effluent (40 μg/L) [1,8], as well as concentrations twice that amount (80 μg/L), cause detectable changes in aggressive behavior over the course of five months exposure time. Including a treatment group exposed to twice the current levels in nature seemed justified given the remarkable rate at which metformin prescriptions are increasing across the US and around the world [3]. The relative ease with which B. splendens normally high levels of aggressive behavior can be observed and quantified when presented with a mirror or other appropriate “releasers” [21] make them an excellent study system for addressing the effects of metformin exposure on aggression in aquatic life.

Materials and methods

Subjects and experimental design

Sexually mature male Siamese fighting fish (B. splendens) purchased from the online breeder AquariumFish.net (https://aquariumfish.net/home.htm) were used to examine how chronic exposure to waterborne metformin influences behavior when presented with a conspecific dummy male stimulus. Male subjects received multiple trials with the dummy stimulus both before and after exposure to metformin. Although the exact genetic background of the subjects is not known, all males were purchased from the same source and were of the same color variety (blue/purple).
The fish were housed singly in 9.5 L (30 X 15 X 15 cm) home tanks where they remained for the five month duration of the study. Each tank was covered on three sides with white paper so that the fish could not see males in neighboring tanks. One long side of the tank was left uncovered for behavioral observations. Males remained isolated in their home tanks for 14 days prior to initiating the behavioral assays in order to reduce the effects that prior social experience might have on their responses.

Subjects were divided randomly into two treatment groups and a control group: fish in the **low-dose treatment** (n = 18) were exposed to 40 μg/L metformin [water borne levels comparable to those found in WWTP effluent;1,8] fish in the **high-dose treatment** (n = 17) were exposed to 80 μg/L metformin; and fish in the **control group** (n = 17) were exposed to an ethanol solution identical to those used to prepare metformin for each of the treatment groups, but containing no trace of the drug (described below). Following a pre-exposure assessment of behavioral responses to a dummy male stimulus (described in detail below), all fish experienced water changes in their home tanks wherein metformin was introduced at the levels described above. Subjects spent four weeks in their metformin-treated home tanks before a second round of identical behavioral tests were conducted. The fish remained in their metformin-treated home tanks for another 16 weeks before a final round of behavioral tests were completed.

**Metformin chemical preparation and water change procedure**

Metformin (1,1-dimethylbiguanidine hydrochloride; CAS # 1115–70–4) and ethanol (200 proof; CAS # 64–17–5) were purchased from Sigma-Aldrich. A 100-mg/L stock solution of metformin was prepared by adding 50 mg of metformin to 10 mL of ethanol, stirring until dissolved, and then adding this to 490 mL of Milli-Q ultrapure water (EMD Millipore). A final metformin concentration of 40 μg/L was achieved by adding 2 mL of stock solution to 5 L of dechlorinated water in the home tanks of each subject designated for the **low-dose treatment**. A final metformin concentration of 80 μg/L was achieved by adding 4 mL of this stock solution to 5 L of dechlorinated water in the home tanks of subjects designated for the **high-dose treatment**. A stock solution for **control** tanks was prepared by adding 10 mL of ethanol to 490 mL of ultrapure water. Four mL of this stock solution was added to 5 L of dechlorinated water in each control tank.

Water in each of the subject tanks was changed once a week for the duration of the study with no external source of aeration or filtration provided. For the first three weeks, water changes involved changing 100% of existing tank water with fresh, dechlorinated tap water from the laboratory. Upon completion of the two-week lab acclimation and pre-exposure behavioral testing phases of the study, the once-per-week water changes proceeded as follows: each water change involved transferring the fish, along with a small amount of their tank water, from their home tanks into temporary holding containers. The home tanks were then drained and refilled with five liters of dechlorinated tap water. For subjects designated as **controls**, 4 mL of ethanol stock solution was added to each home tank. For subjects assigned to the **low-dose treatment**, 2 mL of both ethanol and metformin stock solutions were added to each home tank. Lastly, for subjects assigned to the **high-dose treatment**, 4 mL of metformin stock solution only was added to each home tank. Ethanol was added to the low-dose treatment tanks in order to standardize the total volume of stock solutions (ethanol + metformin) applied to home tanks across all three groups (low-dose, high-dose, and controls). Upon completion of water changes, subject fish were netted and returned to their designated home tanks.

Metformin concentrations in water samples taken from control, low- and high-dose tanks were determined using the method of Majidano and Khuhawar [23]. Mass spectrometry
identified the methlglyoxal derivative at a retention time of 10.9 minutes during the run. Tanks treated to be 0, 40 and 80 μg/L of metformin were sampled repeatedly over the course of five days and measured versus standard solutions of 25, 50 and 100 ppb. Metformin was not detected in samples from control tanks while samples from low- and high-dose tanks kept the desired concentration levels of 40 μg/L and 80 μg/L +/- 10% over the course of seven days of testing. Given these results, we assume there was no significant degradation of metformin in the low- and high-dose treatment tanks between weekly water changes during the course of the five month study period.

**Dummy construction**

In order to control for variation in male behavior among other traits, we used a dummy male as the stimulus. The use of a dummy allowed us to hold all traits such as body size (6.3 cm total length) constant for all behavioral assays, while compromising little in the way of visual detail [24] (Fig 1). The dummy was produced as described in MacLaren et al. [25].

**Behavioral assay procedure**

Subjects receiving the low- and high-dose treatments as well as the control group experienced the same behavioral assay throughout the entire five month study period, including tests completed prior to introducing metformin. The assay involved the presentation of a male dummy at one end of the subject’s home tank. The testing environment consisted of three 9.5 L (30 X

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**Fig 1.** (A) Arrangement of 9.5 L (30 X 15 X 15 cm) aquaria used for all behavioral assays throughout the five-month study period. The subject in its home tank was placed between a pair of aquaria each housing an adjustable motorized belt and pulley system. The male dummy stimulus was attached to the pulley system situated in the aquarium to the right side of the subject’s home tank for all behavioral assays. When the apparatus was turned on, the dummy fish began ‘pacing’ back and forth along the glass facing the subject tank. An iPad was used to record the subject’s behavior for 5 min following the removal of opaque barriers blocking his view of the dummy in the neighboring tank. (B) Image used to create the 6.3 cm total length male dummy used in all behavioral assays.

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15 X 15 cm) aquaria lined up end to end using an apparatus and protocol similar to that used in previous mate preference experiments conducted with poeciliid fishes [26–29]. The subject in its home tank was carefully moved in a manner minimally disruptive to male and placed between a pair of aquaria each housing an adjustable motorized belt and pulley system. The dummy stimulus was attached to the pulley system on the same side of the subject’s home tank for all behavioral assays conducted. Although the pulley system on the opposite side of the tank had no dummy attached, it was turned on whenever the pulley system on the dummy-containing side was engaged in order to control for any effects the system itself might have on subject behavior during testing (Fig 1; see [26] for further details). Following 10 min of acclimation, the apparatus was turned on causing the dummy fish to begin ‘pacing’ back and forth along the glass facing the subject tank. An iPad used to record the subject’s behavior was turned on followed by the removal of opaque barriers used to obstruct the subject’s view of the pulley systems during the 10 min acclimation period. Once the subject male began swimming freely within his home tank, the following behaviors were recorded for a period of 5 min: 1. The number of instances and duration of gill flaring directed towards the male stimulus wherein the subject erects his operculae, fins and tail while facing an opponent; 2. The number of instances and duration of fin spreading directed towards the male dummy in which the subject swims with the side of his body facing the opponent accompanied by erection of fins and tail; and 3. The number of tail beats directed towards the male dummy wherein the subject swims close to a stimulus and flicks his tail in a quick and directional motion towards his ‘opponent’[20–22].

A single ‘sequence’ of behavioral assays consisted of testing each subject, as described above, on three consecutive days. Each subject experienced one sequence of behavioral testing prior to chemical introduction into the subjects’ home tanks followed by two additional sequences of testing after four and 20 weeks exposure time. Four weeks of exposure was chosen in order to compare our results with those of Neimuth et al. [8]. Twenty weeks total exposure time was chosen as a matter of scheduling convenience and to provide time for any effects at the molecular level to potentially impact steroidogenesis and/or behavior.

The research methods presented herein were described in Research Protocol 1RDM1015 and approved by the Institutional Animal Care and Use Committee of Merrimack College. All animal experiments comply with the National Institutes of Health guide for the care and use of Laboratory animals (NIH Publications No. 8023, revised 1978) and such guidelines have been followed.

**Behavioral measures and statistical analyses**

We played back all video data in real time in order to document a total of three dependent behavioral measures including the duration of gill flaring and fin spreading displays, as well as the number of tail beats directed towards the male dummy stimulus per five minute test period. Quantification of all behaviors recorded were made by an observer unaware of the treatment under consideration so as to avoid potential biases from being incorporated into our analyses. The mean scores for each subject over the course of each three day test period were then calculated. A repeated measures analysis of variance (RMANOVA) with post-hoc t-tests were performed on these means to assess if metformin exposure influenced overall behavior to the dummies [30]. Specifically, we compared the mean pre-exposure scores to the mean scores for each of the two subsequent rounds of behavioral assays (4-weeks and 20-weeks exposure) for both the low- and high-dose treatment groups and the control. Three additional sets of RMANOVAs with post-hoc t-tests were conducted comparing the mean scores for low-dose, high-dose, and controls in the pre-exposure, 4-week, and 20-week rounds of testing in
order to assess whether behavioral differences existed among the two treatment groups and the controls within each round. A final set of RMANOVAs were performed on the pre-exposure data for all three groups (low-dose, high-dose, and controls) combined to assess whether subject behavior differed among the three days of testing. In cases where the data to be analyzed violated the assumptions of normality and/or equal variance, the non-parametric Friedman RMANOVA on Ranks was conducted instead [30]. Because each subject was tested three times, all significance levels were adjusted to $p < 0.017$ (Bonferroni procedure [30]). All statistical tests were conducted using the statistical program SigmaPlot Ver. 11.0 (Systat Software, Point Richmond, CA, USA).

**Results**

No trial effects (pre-exposure testing day 1 vs. day 2 vs. day 3) were found for gill flaring display time (Chi-square = 2.138, df = 2; $P = 0.343$), fin spreading display time (Chi-square = 0.122, df = 2; $P = 0.941$), or tail beats (Chi-square = 0.357, df = 2; $P = 0.837$).

**Behavioral effects: Among-treatments comparisons (LD vs HD vs C)**

Compared to the control group, metformin-treated males from both the low-dose (40 μg/L) and high-dose (80 μg/L) treatment groups spent significantly less time exhibiting dummy-directed gill flaring display after both four and 20 weeks of exposure (Table 1; Fig 2). The same result was observed for time spent engaged in fin spreading after 20 weeks, but not after 4 weeks of exposure (Table 1; Fig 3). Similarly, both low- and high-dose subjects directed fewer tail beats to the dummy males after four weeks exposure when compared to controls. However, no significant differences in tail beat frequency were observed among the treatment groups and control after 20-weeks exposure (Table 1; Fig 4).

| Round; Comparison | Behavior       | Mean   | $P$   | C vs. LD | C vs. HD | LD vs. HD |
|-------------------|----------------|--------|-------|----------|----------|-----------|
| PE; C vs LD vs HD | Gill flaring   | 0.98   | 0.385 | ——       | ——       | ——        |
| 4-WK; C vs LD vs HD | Gill flaring   | 17.45  | <0.001$^a$ | T = 3.19 | P = 0.006 | T = 2.71 | P = 0.011 |
| 20-WK; C vs LD vs HD | Gill flaring  | 25.99  | <0.001$^b$ | T = 5.33 | P = <0.001 | T = 1.54 | P = 0.133 |
| PE; C vs LD vs HD | Fin spreading  | 0.60   | 0.553 | ——       | ——       | ——        |
| 4-WK; C vs LD vs HD | Fin spreading  | 4.03   | 0.028 | T = 2.11 | P = 0.083 | T = 2.62 | P = 0.039 |
| 20-WK; C vs LD vs HD | Fin spreading  | 12.25  | <0.001$^c$ | T = 3.25 | P = 0.005 | T = 4.61 | P = 0.117 |
| PE; C vs LD vs HD | Tail beats     | 1.34   | 0.276 | ——       | ——       | ——        |
| 4-WK; C vs LD vs HD | Tail beats     | 6.25   | 0.005$^d$ | T = 2.56 | P = 0.031 | T = 3.30 | P = 0.007 |
| 20-WK; C vs LD vs HD | Tail beats     | 1.47   | 0.246 | ——       | ——       | ——        |

Results of RMANOVAs comparing each subject’s three-day mean behavior scores for gill flaring display duration (sec), fin spreading display duration (sec), and frequency of tail beats among the three treatment groups (i.e. low-dose metformin exposure (LD; n = 18) vs. high dose (HD; n = 17) vs. the controls (C; n = 17)) with associated post-hoc t-tests where applicable (all pairwise multiple comparison procedures; Holm-Sidak method). Separate RMANOVAs were completed on data from each of the three sequences of testing (pre-exposure (PE); 4-weeks of metformin exposure (4-WK); and 20-weeks of metformin exposure (20-WK)). Letter designations (a–d) in association with significant $p$-values (bold) correspond with those presented in Figs 2–4.

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Behavioral effects: Within treatment comparisons (PE vs 4-WK vs 20-WK)

Within the control group, no differences in behavior were observed among the first (pre-exposure), second (4-weeks), or third (20-weeks) sequences of behavioral assays with respect to gill flaring and fin spreading display times, or the frequency of dummy-directed tail beats (Table 2). Neither four nor 20 weeks of metformin exposure at a level of 40 μg/L (low-dose)
affected dummy-directed fin spreading displays or tail beats in comparison with pre-exposure measurements (Table 2). However, low-dose males spent significantly less time engaged in gill flaring after both four and 20 weeks exposure compared to pre-exposure measurements (Table 2; Fig 2).

An effect of metformin exposure at a concentration of 80 μg/L (high-dose) was found for dummy-directed gill flaring and fin spreading displays in comparison with the pre-exposure measurements (Table 2). This effect was observed after both four and 20 weeks of exposure (Fig 4).

Results of RMANOVAs comparing each subject’s three-day mean dummy-directed behavior scores for gill flaring display duration (sec), fin spreading display duration (sec), and frequency of tail beats across three sequences of testing (pre-exposure (PE) vs. 4-weeks of metformin exposure (4-WK) vs. 20-weeks of metformin exposure (20-WK)) with associated post-hoc t-tests where applicable (all pairwise multiple comparison procedures; Holm-Sidak method). Fish in the low-dose treatment group (LD; n = 18) experienced a water borne metformin concentration of 40 μg/L. Fish in the high-dose treatment group (HD; n = 17) experienced a water borne metformin level of 80 μg/L. The control fish (C; n = 17) were never exposed to metformin.

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Table 2. Results of RMANOVAs comparing mean behavior scores across three rounds of testing (PE vs. 4-WK vs. 20-WK).

| Treatment; Comparison | Behavior       | F(df) | P     | PE vs. 20-WK | PE vs. 4-WK | 4-WK vs. 20-WK |
|-----------------------|----------------|-------|-------|--------------|-------------|----------------|
| C; PE vs 4-WK vs 20-WK| Gill flaring   | 1.67  | 0.205 |              |             |                |
| LD; PE vs 4-WK vs 20-WK| Gill flaring   | 12.46| <0.001| T = 4.79     | P = <0.001  | T = 1.18       |
| HD; PE vs 4-WK vs 20-WK| Gill flaring   | 26.14| <0.001| T = 6.51     | P = <0.001  | T = 0.32       |
| C; PE vs 4-WK vs 20-WK| Fin spreading  | 0.29  | 0.747 |              |             |                |
| LD; PE vs 4-WK vs 20-WK| Fin spreading  | 2.70  | 0.082 |              |             |                |
| HD; PE vs 4-WK vs 20-WK| Fin spreading  | 13.41| <0.001| T = 5.14     | P = <0.001  | T = 1.99       |
| C; PE vs 4-WK vs 20-WK| Tail beats     | 2.04  | 0.147 |              |             |                |
| LD; PE vs 4-WK vs 20-WK| Tail beats     | 0.08  | 0.923 |              |             |                |
| HD; PE vs 4-WK vs 20-WK| Tail beats     | 0.90  | 0.415 |              |             |                |

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with males spending significantly less time engaged in these behaviors when compared to pre-exposure levels (Table 2; Figs 2 and 3). However, the number of dummy-directed tail beats observed in the high-dose treatment group did not differ among the pre-exposure, 4-week, or 20-week rounds of testing (Table 2).

Discussion

Our results show that metformin at concentrations in the range of what has been documented in freshwater systems and waste-water effluent (40 μg/L) affects aggressive behavior in *B. splendens*. Metformin-exposed males exhibited less aggression toward the male conspecific stimulus after four weeks when compared to levels measured immediately prior to their exposure, and in comparison to a separate cohort of un-exposed control fish. This effect persisted after 20 weeks exposure as well. Specifically, metformin affected the longer-range displays of gill flaring and fin spreading more so than the frequency of tail beating behavior. Subjects exposed to metformin at a concentration twice that currently observed in nature (80 μg/L) exhibited even more substantial reductions in gill flaring and fin spreading behaviors compared to controls and pre-exposure measurements than those observed in the low-dose treatment group. However, similar to the low-dose subjects, tail beat frequency was affected by 80 μg/L metformin after four weeks, but not 20 weeks of exposure.

Some possible explanations for the differences observed in metformin’s effects on long-range displays vs. tail beats include the fact that tail beating is a less frequently observed behavior than gill flaring and fin spreading displays. Moreover, tail beats typically occur only after fights have escalated and when males have perceived their opponents as a serious threat [19]. Subjects in the present study were unable to make physical contact with the male stimulus during behavioral tests, which might have affected the frequency of tail beats since there was no opportunity for fights to escalate. Alternatively, it could be the proximate physiological basis of tail beating behavior is different enough from those underlying the expression of frontal and broadside displays that it is less affected by metformin exposure.

Our data suggest metformin’s effect on aggressive behavior in *B. splendens* increases with length of exposure, as the frequency of gill flaring and fin spreading both trended downward over the course of the study period (Figs 2 and 3). This trend could be a consequence of habituation to the dummy stimulus and/or effects associated with prolonged social isolation between each round of testing. However, were this the case, we would expect a similar decline in aggressive behavior from the control fish, which was not observed.

Aggression is vital for territory defense and mate acquisition in *B. splendens*. Males build large nests made of bubbles preceding any mating or courtship—a lengthy process that involves taking air from the surface, coating it with sticky mucus, and placing the coated bubbles in a singular large group [31]. The territory selected for nest building serves as the location of all courtship and hatching of the fry and is therefore critical to the male’s reproductive success. In the wild, the nest is essential for egg production. The male places the female’s eggs, once laid, into the nest, which he then maintains and defends until hatching [31]. Any disturbance to these tasks could alter offspring success and have population-level consequences. Changes in male aggressive behavior can therefore have a significant impact on offspring surviving to adulthood. Moreover, male aggression is critical to their survival and successful courtship, even before initiating the nest-building process [22]. When a male’s ability to expel rivals from his territory is compromised as a consequence of a reduction in aggressive behavior, there will most certainly be fitness consequences.

Males respond differently to conspecifics when they have nests versus when they do not [22], which could serve as a confounding variable in our analysis of the results of this study.
However, all subjects from both treatment groups and the control showed evidence of bubble nest building behavior at various points in time during the course of the present study with no obvious differences among them (personal observation). We therefore assume that differences in nest building behavior among males cannot account for the differences in aggression levels observed among the LD, HD and control groups throughout the study period. That said, the potential effects of metformin on nest building behavior of male *B. splendens* deserves more careful and thorough investigation.

The present study does not experimentally address any potential proximate bases for metformin’s effect on aggressive behavior in *B. splendens*. However, hormonal changes in association with metformin exposure serve as an excellent starting point for such an investigation since hormonal differences are often associated with behavioral differences. For example, elevated levels of serotonin in fish are observed in more subordinate individuals, leading to reduced fighting behavior [32,33], while glucocorticoid stress hormones and testosterone levels affect the willingness to engage in and the probability of winning fights with rival conspecifics [34]. We speculate that waterborne exposure to metformin may affect aggressive behavior in *B. splendens* by altering steroid hormone production in males. Metformin has been implicated as a potential EDC in fathead minnows (*Pimephales promelas*), inducing vitellogenin mRNA expression in the livers of exposed males [8]—an early and sensitive indicator of exposure to EDCs [35–37]. The fathead minnow study [8] did not involve assessment of the effects of metformin on behavior specifically, and their results provided no evidence of hormone or reproductive changes. However, the authors argue this could be because of the length or timing of exposure—four weeks at 40 μg/L—a dosage and exposure time examined in the low-dose treatment of the present study. Whether or not a consequence of endocrine disruption, it is clear that 40 μg/L metformin exposure affects aggressive behavior in *B. splendens* after four weeks, with longer exposure time (20 weeks) facilitating even more substantial behavioral changes when compared with unexposed control males.

It could be that metformin has a negligible impact on hormone production, secretion rates and/or their target cells throughout the body. Metformin’s biochemical structure in fact does not resemble those of hormone-like compounds typically identified as endocrine disruptors. Rather, metformin may be affecting aggressive behavior via an entirely different as yet unidentified mechanism. The dearth of studies addressing how non-endocrine disrupting chemicals may interfere with normal behavior make the results of the present study all the more impactful in addressing pharmaceuticals and their effects on aquatic organisms in nature.

Lastly, it is important to note that such static and constant exposures as used in the present study differ from what aquatic organisms would normally receive in a more natural setting. Chemicals from WWTP are dispersed in a manner where organisms will never experience constant levels. Moreover, aquatic life are exposed to a variety of chemicals simultaneously, which may affect the impact any one chemical might otherwise have on its own. Clearly there is much work to be done in order for science to get a better handle on the potential health effects associated with exposure to metformin (among many other pharmaceuticals) as it occurs in nature.

In summary, overall levels of aggression were reduced by chronic waterborne exposure to metformin. These results have evolutionary fitness implications for species where aggression is vital for territorial defense and mate acquisition. The present study demonstrates the need for further examination of the drug’s impact on aquatic life. Studies of the effects of metformin at a wider range of doses, over longer periods of exposure, and in other species, should offer a better understanding of the endocrine and behavioral impacts of metformin exposure and its effects on survival and reproductive success in the aquatic environment.
Supporting information

S1 Table. Metformin Betta raw data D1 thru D3 for PLOS ONE. Raw behavioral data collected and used in all statistical analyses conducted in the present study.
(XLSX)

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