Assessment of the Possible Association of Air Pollutants PM$_{10}$, O$_3$, NO$_2$ With an Increase in Cardiovascular, Respiratory, and Diabetes Mortality in Panama City

A 2003 to 2013 Data Analysis

Julio Zúñiga, MD, Musharaf Tarajia, MD, Víctor Herrera, MSc, Wilfredo Urriola, MSc, Beatriz Gómez, MSc, and Jorge Motta, MD

Abstract: In recent years, Panama has experienced a marked economic growth, and this, in turn, has been associated with rapid urban development and degradation of air quality. This study is the first evaluation done in Panama on the association between air pollution and mortality. Our objective was to assess the possible association between monthly levels of PM$_{10}$, O$_3$, and NO$_2$ and cardiovascular, respiratory, and diabetes mortality, as well as the seasonal variation of mortality in Panama City, Panama.

The study was conducted in Panama City, using air pollution data from January 2003 to December 2013. We utilized a Poisson regression model based on generalized linear models, to evaluate the association between PM$_{10}$, NO$_2$, and O$_3$ exposure and mortality from diabetes, cardiovascular, and respiratory diseases. The sample size for PM$_{10}$, NO$_2$, and O$_3$ was 132, 132, and 108 monthly averages, respectively.

We found that levels of PM$_{10}$, O$_3$, and NO$_2$ were associated with increases in cardiovascular, respiratory, and diabetes mortality. For PM$_{10}$ levels $\geq 40$ µg/m$^3$, we found an increase in cardiovascular mortality of 9.7% (CI 5.8–13.6%), and an increase of 12.6% (CI 0.2–24.2%) in respiratory mortality. For O$_3$ levels $\geq 20$ µg/m$^3$ we found an increase of 32.4% (CI 14.6–52.9) in respiratory mortality, after a 2-month lag period following exposure in the 65 to $<74$ year-old age group. For NO$_2$ levels $\geq 20$ µg/m$^3$ we found an increase in respiratory mortality of 11.2% (CI 1.9–21.3), after a 2-month lag period following exposure among those aged between 65 and $<74$ years.

There could be an association between the air pollution in Panama City and an increase in cardiovascular, respiratory, and diabetes mortality. This study confirms the urgent need to improve the measurement frequency of air pollutants in Panama.

INTRODUCTION

Air pollution is an important public health problem that causes adverse health and economic consequences. In 2012, a report by the World Health Organization (WHO) determined that 3.7 million people died as a result of air pollution, especially in rapidly growing developing countries. Among the air pollutants, particulate matter with an aerodynamic diameter of $<10$ µm (PM$_{10}$) is mostly associated with the air pollution in large cities. Nitrogen dioxide (NO$_2$) and sulfur dioxide (SO$_2$) are the primary air pollutants associated with vehicles. On the other hand, levels of ozone (O$_3$) have been increasingly associated with climate change. According to the WHO, increased levels of these air pollutants lead to multiple adverse health effects.

Epidemiology studies, primarily in Asia, have shown adverse health effects of short- and long-term exposure to air pollutants are associated not only with morbidity manifested by myocardial infarction or stroke, but also with mortality associated to diabetes, respiratory, and cardiovascular disease. In Latin America, published data related to the health effects of air pollutants are limited to Mexico, Brazil, and Chile. Based on these findings, many countries have identified air pollution as a public health problem because of the multiple evidence of its effect on population health and mortality. Some countries have gone further, implementing public measures to control air pollutants, with good results.

According to the World Bank, Panama has had one of the highest rates of economic growth in Latin America in the last decade, which has been associated with rapid urban growth especially in Panama City. This rapid urban growth, in turn, has been linked to increased levels of air pollutants like particulate matter with an aerodynamic diameter of $<2.5$ µm (PM$_{2.5}$), reaching on some days concentrations $>25$ µg/m$^3$. Although levels of some air pollutants have been recorded for several years, there have been no published reports analyzing the association of air pollution with mortality in Panama or in Central America.

Abbreviations: ETESA = Electric Transmission Company, ICD-10 = 10th International Classification of Disease, IQR = interquartile range, NO$_2$ = nitrogen dioxide, O$_3$ = ozone, PM$_{10}$ = particulate matter $\leq10$ µm, PM$_{2.5}$ = particulate matter $\leq2.5$ µm, SI = seasonal index, WHO = World Health Organization, WMO = World Meteorological Organization.
The purpose of this study is to assess the possible association between monthly levels of PM10, O3, and NO2 and cardiovascular, respiratory, and diabetes mortality and also the seasonal variation of mortality in Panama City.

METHODS

Study Area
Panama City is Panama’s largest city, with a total surface area of 280 km², inhabited by ~1.2 million people, which represent ~40% of the total population of the country. Panama City is composed of the district of Panama and, partially, by the district of San Miguelito. Both districts account for 16 subdivisions (corregimientos).

Procedures
We analyzed air pollution data of NO2 and PM10 from January 2003 to December 2013 and O3 data from January 2003 to December 2011 in Panama City by using a Poisson regression model. Data for O3 was not available for 2012 and 2013; therefore, these years were not included in the analyzed model. We could only use this time range for analysis because limited data on air pollutants were available before 2003. Also, the study only included Panama City because there were no air pollution data and lack of high-quality mortality data in other parts of the country. In this study, we also report a seasonality descriptive pattern by using a time series decomposition procedure.

The sample size consisted of a 132 monthly average measurements for NO2 and PM10 and 108 monthly measurements for O3. Two daily measurements, at 8AM and at 4PM, were averaged for a daily value. This process was performed twice a week. Then, all the measurements available in each month were averaged for a single monthly value. For NO2 and PM10 twice a week values were available. For O3, once or twice a week values were averaged (see Weather and Air Pollution Data section).

DATA COLLECTION

Mortality
Monthly mortality statistics from January 2003 to December 2013 were obtained from the National Institute of Statistics and Census. Panama City has the highest quality of mortality data in the country, were a qualified doctor certifies all deaths.

The causes of death were coded according to the 10th International Classification of Diseases (ICD-10) and segregated into 3 categories: cardiovascular mortality (I00–I99), respiratory mortality (J00–J99), and diabetes mortality (E10–E14). We excluded Influenza (J10–J11) of respiratory mortality, as it was considered a potential confounding variable in the model. The Ministry of Health of Panama provided the data of influenza cases. Mortality was stratified according to 4 age groups: <65 years, 65 years to 74 years, 75 to 84 years, and ≥85 years.

Weather and Air Pollution Data
For the study period, we obtained daily maximum and minimum temperatures taken between 12:00 h and 00:00 h in degrees Celsius (°C), percent relative humidity (%), precipitation in millimeters (mm), wind speed at 10 m/s and calculated a monthly average for these weather variables. The Electric Transmission Company (ETESA) supplied all the weather data (without missing values). The temperature was obtained from the station of Hato Pintado #142–020, 45 meter elevation, 9°00’33"N latitude, 79°30’52"W longitude and the rest of weather data from the station of Tocumen #14402 WMO (World Meteorological Organization) # 78792, 14 meter elevation, 9°03’00”N latitude, longitude 79°22’00”W longitude.

We utilized 5 air pollution-monitoring stations distributed in different representative locations of the city. For PM10 and NO2 data were available twice a week (Monday and Thursday, or Tuesday and Friday) from the different air pollution-monitoring station. Forty values (8 per station by 5 stations) for PM10 and NO2 each month were averaged to produce a single monthly value, using a method similar to another study. The Institute of Specialized Analysis of the University of Panama, which provided the air pollution data, measured O3 once or twice per week; therefore, 20 to 40 daily measurements per month were average to obtain a representative single monthly value (see Figure 1).

The passive tubes technique was used for the measurement of O3 and NO2, and a Harvard-type impactor was used for the measurement of PM10. All the stations used the same standards and measurement protocols. In 6 instances, there was a temporal interruption of a daily measurement; therefore we estimated the absent daily measurement by using the monthly mean of the contaminant of the pollutant-monitoring station where the value was not recorded.

SEASONALITY
We calculated the seasonality of monthly mortality and pollutant levels for the period January 2003 to December 2013 by utilizing cardiovascular, respiratory, and diabetes mortality, and the levels of PM10, O3, and NO2. The seasonal component was calculated by the moving averages method, using data 6 months before and 6 month after the calculated month. We assumed that the pattern of variation remained constant from year to year (random error equal to zero) and that it could be quantified with index numbers (seasonal index) when the aggregation scheme was multiplicative. The seasonal index (SI) was used to summarize the behavior of the contaminants and disease variable per month.

Single Pollutants Model
For the study period, a Poisson regression model, based on generalized linear models, was used to assess changes in mortality risk for cardiovascular, respiratory, and diabetes diseases and the association to the levels of O3, PM10, and NO2. The dependent variable was the monthly and total mortality stratified by the following age groups: <65 years, 65 to 74 years, 75 to 84 years, and ≥85 years. For the gender analysis we did not find an association between air pollution and mortality. The significance level was established at 0.05.

We established 2 dummy variables for the analysis of each pollutant (PM10≥40 μg/m³, O3 ≥20 μg/m³, NO2 ≥20 μg/m³) to determine the point considered to be the highest level of contamination.

We used a cut-off value of 20 μg/m³ for O3 as suggested by a meta-analysis result. For PM10 (≥40 μg/m³) and NO2 (≥20 μg/m³) we used the cutoff values using a methodology similar to other reports. Briefly, for PM10 and NO2, we looked for the air contaminant value in which we found an effect with an adequate sample size. These values were higher than others used previously. Lower cutoff values could not
be used as lower values of PM10 and NO2 were scarce to perform the analysis. The validity of each model was evaluated by the omnibus test for logistic regression and by the likelihood ratio test. Each model was controlled for relative humidity, precipitation, 10-meter wind speed, mean temperature, linear trend in mortality, and influenza cases (see equation in supplementary data, http://links.lww.com/MD/A634). We used the influenza variable as a reference point (odds ratio: 1.00). The effect of environmental pollutants on mortality was also evaluated using 1- and 2-month mortality lags. The results are presented as a percentage change in mortality with their respective 95% confidence intervals (CI).

Spearman’s rank correlation coefficient analysis was performed using monthly values of PM10, O3, NO2, and relative humidity, average temperature, and wind speed, in order to assess the degree of correlation between these variables. Data were analyzed using SPSS 19.0 Statistical Package (IBM, Armonk, NY) and plotted in GraphPad Prism 6 (La Joya, CA). An ethics committee reviewed and approved the study.

RESULTS

Mortality and Air Pollution Data
From 2003 to 2013, there were 18,468 cardiovascular, 5709 respiratory, and 4404 deaths due to diabetes in the 16 subdivisions (corregimientos) of Panama City. Table 1 summarizes the mean monthly mortality, air pollution levels, average temperature, relative humidity, and wind speed in the city. The mean monthly mortality attributed to cardiovascular and respiratory diseases and diabetes was 139.9 deaths/month, 43.2 deaths/month, and 33.4 deaths/month, respectively. The mean monthly concentrations for PM10, O3, and NO2 were 13.8 μg/m3, 43.8 μg/m3, and 28.9 μg/m3 with interquartile ranges (IQR) of 5.97 μg/m3, 11.28 μg/m3, and 5.82 μg/m3, respectively. The mean temperature was 27.9 °C with an IQR of 1.1 °C. The average monthly relative humidity was 74.9% with an IQR of 11 and the mean 10-m wind speed was 1.82 m/s with an IQR of 0.64 m/s.

Seasonality
Figure 2 summarizes mortality and pollution SI. The lowest SI of cardiovascular disease (SI 88) and diabetes mortality (SI 76) occurred during February. The peak of mortality for cardiovascular disease occurred in June (SI 108) and for diabetes in September (SI 110). As for respiratory mortality, the lowest point occurred in April (SI 83) the highest value occurred in June (SI 120). O3 pollution presented its highest SI in March (SI 119) and its lowest SI in August (SI 96), whereas PM10 registered its highest SI in April (SI 109) and its lowest SI in October (SI 88). NO2 reached its highest SI level in February (SI 113), but there was less variation in NO2 levels during the year.
Results of the Model

Figure 3 shows the subacute association between total mortality and air pollution (PM10, O3, and NO2) from 2003 to 2013 in Panama City, stratified by age and by 1 to 2 month interval lag periods (see Table 1 in supplementary file, http://links.lww.com/MD/A634). For PM10 ≥ 40 μg/m³, the study revealed an increase of 9.7% (CI 5.8% to 13.6%) in total cardiovascular mortality and of 5.8% (CI 1.9–9.7%) 1 month after the time of exposure and we did not obtain a statistically significant association 2 months after the exposure. Regarding respiratory mortality, there was an increase of 12.6% (CI 0.2–24.2%) in total mortality, a 5.6% (CI 0–12.3%) increase after a 1-month lag period, and an 11.2% (CI 0.7–22.8%) increase after a 2-month lag period from the time exposure. When stratified by age, respiratory mortality increased to 30% (CI 7.1–58.0%) in the <65 year-old age group. For total diabetes mortality and after a 1-month lag period. For total respiratory mortality, we did not obtain a statistically significant association without lag period and after a 1-month lag period and a 2.3% increase in mortality after a 1-month lag period.

Levels of NO2 ≥20 μg/m³ were associated with an increase in cardiovascular mortality of 2.3% (CI 0.1–8.2%) in the ≥85 year-old group and a 6.7% (CI 0.9–12.8%) increase after a 1-month lag period. For total respiratory mortality, we did not obtain a statistically significant association without lag period and after a 1-month lag period, although we observed an increase of 11.2% (CI 1.9–21.3%) after a 2-month lag period. In the ≤65 year age group we observed a 40.7% (CI 18.8–66.5%) increase after a 2-month lag period from the time of exposure. NO2 was not significantly associated with an effect on total diabetes mortality, after and without lag periods.

Table 2 shows a Spearman correlation analysis between environmental pollutants and weather variables. We observed that environmental pollutants had an inverse, but not statistically significant (P>0.05), correlation with wind speed. The levels of O3 and PM10 were inversely correlated with relative humidity levels. The levels of PM10 were inversely correlated with relative humidity (r = 0.363, P = 0.03). O3 and NO2 were also inversely correlated (r = −0.267, P = 0.005).
DISCUSSION

This is the first study done in Panama and Central America that has made an assessment of the possible association between air pollutants and a subacute increase in cardiovascular, respiratory, and diabetes mortality. Using statistical models that assess the subacute individual impact of each pollutant, we conclude that PM$_{10}$, O$_3$, and NO$_2$ represents another factor together with comorbidities, treatment adherence, age, and socioeconomic status that could be increasing mortality statistics in Panama City, Panama.

The pathophysiologic effects of air pollutants have been evaluated in several controlled animal studies and, because of ethical considerations, in a few human studies. Studies have shown that inhalation of particulate pollutants cause an inflammatory response of the airways and of the endothelium, associated to neutrophil and monocytes migration, a rise in the levels of inflammatory cytokines such as TNF-$\alpha$, IL-1, and IL-6, and an increase in oxidative stress.$^{32-36}$ Rats exposed to concentrated ambient pollutants have shown a predisposition to longer inflammatory effects of their respiratory airways.$^{37}$ In humans, hemodynamic studies have shown significant acute blood pressure increases after exposure to air pollutant particles or to controlled levels of O$_3$. $^{38-40}$

In patients with diabetes due to own endothelial alteration of the disease, the inflammatory response product of air pollution and cardiovascular effect is greater, predisposing to the development of cardiovascular events and inflammation of the

FIGURE 2. Monthly mortality and air pollutants seasonality in Panama City from 2003 to 2013. CV = cardiovascular mortality, NO$_2$ = nitrogen dioxide, O$_3$ = ozone, this value only was registered in 2003 to 2011, PM$_{10}$ = particulate matter $\leq$ 10 $\mu$g.

FIGURE 3. Percentage change (mean and 95%CI) in mortality associated with an increase in $\geq$ 40 $\mu$g/m$^3$ PM$_{10}$ (monthly lag 0, 1, 2), $\geq$ 20 $\mu$g/m$^3$ in O$_3$ (monthly lag 0, 1, 2), and $\geq$ 20 $\mu$g/m$^3$ in NO$_2$ (monthly lag 0, 1, 2) in Panama City from 2003 to 2013. The association with ozone only was analyzed from 2003 to 2011. Supplemental Digital Content. Odds ratio values (mean and 95%CI) in mortality associated with an increase in $\geq$ 40 $\mu$g/m$^3$ PM$_{10}$ (lag 0, 1, 2), $\geq$ 20 $\mu$g/m$^3$ in O$_3$ (lag 0, 1, 2), and $\geq$ 20 $\mu$g/m$^3$ in NO$_2$ (lag 0, 1, 2) in Panama City from 2003 to 2013. CI = confidence interval, NO$_2$ = nitrogen dioxide, PM$_{10}$ = particulate matter $\leq$ 10 $\mu$m.
respiratory epithelium, which seems to be more evident in patients <65 without adequate control. Other studies have shown an increased risk of developing a prothrombotic state after exposure to high levels of airborne particulate contaminants, which is more marked in diabetics.

PM$_{10}$

As described in this study, many investigations have demonstrated an association between the short- and long-term exposure to high levels of PM$_{10}$ and an increased risk of cardiovascular and respiratory mortality. A study done in 2 Austrian cities, utilizing a similar statistical analysis as in this study, showed an increase mortality risk from cardiovascular disease of 2.0% (95% CI, 0.9–3.1%) and an increase mortality risk from respiratory disease of 3.0% (95% CI, 0.5–5.5%) after a 15-days of exposure to PM$_{10}$.

The effect of the air pollutants is more evident in respiratory and cardiovascular mortality without lag period and after 1 month lag period. The <65 years and ≥85 year-old groups appear to be more vulnerable. This may be due to a combination effect of air pollution and other factors including socioeconomic status, comorbidities, lifestyle, and treatment adherence. We did not observe a statistically significant association when exploring a 2-month lag period.

When compared to other studies, the marked increase risk in respiratory mortality due to PM$_{10}$ noted in our study (12.6%) could be explained by the exposure to high levels of PM$_{10}$ (≥40 μg/m$^3$) registered not frequently in the city and the increased exposure time.

In our study, as well as in another report, the association of PM$_{10}$ to respiratory mortality was more pronounced in <65 years and ≥85 year-old groups, which because of age-related changes, have a higher prevalence of respiratory pathologies that might be worsened by exposure to this pollutant.

For diabetes mortality we did not find a consistent association with air pollutants. Some experimental studies suggest increases in prothrombotic states that could be associated with mortality events, but more epidemiological studies are needed to conclude the subacute effect of PM$_{10}$ in diabetes mortality.

O$_3$

Contrary to other studies that did not find an association between exposure to O$_3$ and mortality, our study found a significant association between levels of O$_3$ ≥20 μg/m$^3$ and mortality from total cardiovascular mortality (6.9%) and total respiratory mortality (14.2%) disease. The association of O$_3$ exposure to respiratory mortality was observed after 1- and 2-month lag periods in the ≥75 to 84 age group. This relation could suggest that the subacute exposure to O$_3$ has a possible effect, which is more evident in older age groups, a study that evaluate the acute exposure of O$_3$ in older age groups find an increase in cardiovascular and respiratory mortality. This finding is in line with the acute and subacute injury seen in experimental studies that have evaluated the effects of O$_3$ exposure on the respiratory and cardiovascular systems of rats and in human tissue culture studies.

NO$_2$

Research studies assessing the short- and long-term effects of NO$_2$ on human health have shown an increased risk of cardiovascular, respiratory, and diabetes mortality for NO$_2$ levels ≥10 μg/m$^3$. One study also demonstrated, after subacute exposure of NO$_2$, an increase in mortality of 4.6% (95% CI, 0.1–4.1%) from cardiovascular disease and an increase in mortality of 6.7% (95% CI, 2.7–10.8%) from respiratory mortality.

Our study as have the above-mentioned studies, found an association between increased levels of NO$_2$ and cardiovascular mortality but only in the ≥74 year old group, as well as an increase in respiratory mortality only in the ≥65 to 74-year-old group. Furthermore, in both groups, the increase in mortality was shown after exposure to NO$_2$ after a 2-month lag period. We believe the difference found in our study compared to others, concerning the health effects of NO$_2$, might be due to the sampling frequency of NO$_2$ in our study. A more accurate assessment of the relation of NO$_2$ exposure to mortality could have been achieved with daily measurements.

Seasonal mortality patterns have been studied in several countries through various analytical models. Some studies have established an association of seasonal mortality with weather variables such as temperature and wind speed whereas others have demonstrated the role of wind speed, wind direction, and other weather variables on the levels of air pollutants. Our study revealed a seasonal mortality pattern characterized by a lower mortality SI for cardiovascular disease and diabetes during January to March. We hypothesized that the northeast trade winds, that are more prominently present during the months of January and February (early dry season of Panama), could have a cleansing effect of air pollutants of the city and that this wind effect could have played a role in the observed reduction of mortality during those months. Although our analysis showed a slight inverse relation between the levels of air pollutants and wind speed, this inverse relation did not reach statistical significance. NO$_2$ and O$_3$ that are intimately linked through atmospheric chemistry and continuous interchange over very short timescales. In our study we observed a slightly negative correlation (P < 0.05). This correlation should be addressed in other studies.

Our study was limited by the sampling frequency of air pollution data and by the availability only of monthly mortality data, which was used to analyze the possible

### TABLE 2. Spearman Correlation Coefficients Between Air Pollutants and Weather Variables. The P Values Are Represented in Parentheses After Each Correlation Coefficient

| Pollutant | O$_3$ (P) | NO$_2$ (P) | Mean Temperature (P) | %Relative Humidity (P) | Wind Speed (m/seg) (P) |
|-----------|-----------|------------|----------------------|------------------------|-----------------------|
| PM$_{10}$ | 0.062 (0.527) | 0.074 (0.400) | 0.007 (0.939) | −0.363 (0.03) | −0.042 (0.635) |
| O$_3$     | −0.267 (0.005) | 0.105 (0.281) | −0.034 (0.724) | −0.124 (0.202) | |
| NO$_2$    | −0.069 (0.447) | 0.196 (0.031) | −0.046 (0.600) | |

NO$_2$ = nitrogen dioxide, O$_3$ = ozone, PM$_{10}$ = particulate matter ≤10 μm.
association between air pollution and mortality in Panama City. Daily sampling of pollutants and daily mortality data surely would have given us a more precise assessment of the relation between mortality from cardiovascular disease, respiratory disease, and diabetes and air pollution. The other limitation was the frequency of weekly sampling and the year period of measurements of O₃ levels, which extended from 2003 to 2011, whereas the period of measurement of the other contaminants was from 2003 to 2013. This reduced the sample size in the analysis might have decreased the capacity to determine more precisely the deleterious effect of O₃ shown in the study. Also, the irregular time of observation of O₃ and the other 2 pollutants could have affected the correlation analysis between environmental pollutants and weather variables.

Panama is a country that has sustained high economic growth and a rapid urban expansion in recent years, and it needs better monitoring of air pollutant levels, its health impacts, and the control of its sources. This study showed a possible deleterious association between the air pollution in Panama City and cardiovascular, respiratory, and diabetes mortality. Also, it confirms the urgent need to improve the sampling frequency of air pollutants in the country. Finally, we believe that this will allow us to precisely address the health effects, with the aim of developing policies to protect the citizens.

ACKNOWLEDGMENTS

The authors thank the Institute of Specialized Analysis of the University of Panama, the meteorological department of the Electric Transmission Company (ETESA), the national Comp-troller’s Office (Contraloría General de la República de Panamá), and the Ministry of Health of Panama for providing data on air pollution, weather conditions, mortality, and cases of influenza, respectively. Finally, they would like to thank MSc. Eva Castillero for supporting us in the preparation of the map of Panama City.

REFERENCES

1. Gao M, Guttikunda SK, Carmichael GR, et al. Health impacts and economic losses assessment of the 2013 severe haze event in Beijing area. Sci Total Environ. 2015;511:553–561.
2. Haucke F, Bruckner U. First approaches to the monetary impact of environmental health disturbances in Germany. Health Policy. 2010;94:34–44.
3. Leem JH, Kim ST, Kim HC. Public-health impact of outdoor air pollution for 2(nd) air pollution management policy in Seoul metropolitan area, Korea. Ann Occup Environ Med. 2015;27:7.
4. Ambient (outdoor) Air Quality and Health, Fact sheet, World Health Organization, 2014.
5. Brunekreef B, Beelen R, Hoek G, et al. Effects of long-term exposure to traffic-related air pollution on respiratory and cardiovascular mortality in the Netherlands: the NLCS-AIR study. Res Rep Health Eff Inst. 2009;5:71; discussion 73-89.
6. Tang X, Wilson SR, Solomon KR, et al. Changes in air quality and tropospheric composition due to depletion of stratospheric ozone and interactions with climate. Photochem Photobiol Sci. 2011;10:280–291.
7. Air Quality Guidelines for Particulate Matter, Ozone, Nitrogen Dioxide and Sulfur Dioxide, World Health Organization, 2005.
8. Biggeri A, Bellini P, Terracini B. Meta-analysis of the Italian studies on short-term effects of air pollution. Epidemiol Prev. 2001;25(2 Suppl):1–71.
9. Brook RD, Cakmak S, Turner MC, et al. Long-term fine particulate matter exposure and mortality from diabetes in Canada. Diabetes Care. 2013;36:3313–3320.
10. Chen H, Goldberg MS, Villeneuve PJ. A systematic review of the relation between long-term exposure to ambient air pollution and chronic diseases. Rev Environ Health. 2008;23:243–297.
11. Jerrett M, Burnett RT, Pope CA, et al. Long-term ozone exposure and mortality. N Engl J Med. 2009;360:1085–1095.
12. Wong CM, Vichit-Vadakan N, Vajanapoom N, et al. Part 5. Public health and air pollution in Asia (PAPA): a combined analysis of four studies of air pollution and mortality. Res Rep Health Eff Inst. 2010;377–418.
13. Hoffmann B, Weinmayr G, Hennig F, et al. Air quality, stroke, and coronary events. Desch Arztlbl Int. 2015;112:195–201.
14. Romieu I, Gouveia N, Cifuentes LA, et al. Multicity study of air pollution and mortality in Latin America (the ESCALAN study). Res Rep Health Eff Inst. 2012;171:5–86.
15. Fischer PH, Marra M, Ameling CB, et al. Air pollution and mortality in seven million adults: The Dutch Environmental Long-itudinal Study (DUELS). Environ Health Perspect. 2015;123:697–704.
16. Lin J-H, Yen T-H, Weng C-H, et al. Environmental NO₂ level is associated with 2-year mortality in patients undergoing peritoneal dialysis. Medicine. 2015;94:e368.
17. Madrigano J, Jack D, Anderson GB, et al. Temperature, ozone, and mortality in urban and non-urban counties in the northeastern United States. Environ Health. 2015;14:3.
18. Anenberg SC, Schwartz J, Shindell D, et al. Global air quality and health co-benefits of mitigating near-term climate change through methane and black carbon emission controls. Environ Health Perspect. 2012;120:831–839.
19. Chanel O, Henschel S, Goodman PG, et al. Economic valuation of the mortality benefits of a regulation on SO₂ in 20 European cities. Eur J Public Health. 2014;24:631–637.
20. Broome RA, Fann N, Cristina TJ, et al. The health benefits of reducing air pollution in Sydney, Australia. Environ Res. 2015;143(Pt 1 A):19–25.
21. ITINoAo Sciences, Urban Water Challenge in the Americas: A Perspective from the Academies of Science, México DF: IANAS and UNESCO, volume 1, 2015.
22. Wilfredo Urriola García HGC, Almer E, Vásquez. Determination of PM 2.5 particulate matter and air quality in the digital station located at Campus Harnmodio Arias Madrid in the city of Panamá. J Environ Public Health Panama. 2014;2:1–10.
23. Contraloría General de la República: Instituto nacional de Estadística y Censo. Panama Republic, census 2010. INEC, Vol 3. Obtained on January 15, 2015 http://www.contraloria.gob.pa/inec/Redatam/censosmpa.
24. Mathers CD, Fat DM, Inoue M, et al. Counting the dead and what they died from: an assessment of the global status of cause of death data. Bull World Health Organ. 2005;83:171–177.
25. Ou CQ, Song YF, Yang J, et al. Excess winter mortality and cold temperatures in a subtropical city, Guangzhou, China. PLoS One. 2013;8:e77150.
26. De Santis F, Dogeroglu T, Menichelli S, et al. The use of a new passive sampler for ozone and nitrogen oxides monitoring in ecological effects research. Scientific World J. 2001;1:475–482.
27. Philip Demokritou SJL, Stephen T. Ferguson A compact multistage (cascade) impactor for the characterization of atmospheric aerosols. Aerosol Sci. 2004;35:281–299.
28. Brunekreef B, Holgate ST. Air pollution and health. Lancet. 2002;360:1233–1242.
29. Bell ML, Dominici F, Samet JM. A meta-analysis of time-series studies of ozone and mortality with comparison to the national morbidity, mortality, and air pollution study. *Epidemiology* 2005;16:436–445.

30. Daniels MJ, Dominici F, Zeger SL, et al. The National Morbidity, Mortality, and Air Pollution Study. Part III: PM10 concentration–response curves and thresholds for the 20 largest US cities. *Res Rep Health Eff Inst*. 2004;94 (P3):1–21; discussion 23-30.

31. Li C, Fang D, Xu D, et al. Main air pollutants and diabetes-associated mortality: a systematic review and meta-analysis. *Eur J Endocrinol*. 2014;171:R183–190.

32. Hajat A, Allison M, Diez-Roux AV, et al. Long-term exposure to air pollution and markers of inflammation, coagulation, and endothelial activation: a repeat-measures analysis in the Multi-Ethnic Study of Atherosclerosis (MESA). *Epidemiology*. 2015;26:310–320.

33. Ghio AJ, Smith CB, Madden MC. Diesel exhaust particles and airway inflammation. *Curr Opin Pulm Med*. 2012;18:144–150.

34. Imrich A, Ning Y, Lawrence J, et al. Alveolar macrophage cytokine response to air pollution particles: oxidant mechanisms. *Toxicol Appl Pharmacol*. 2007;218:256–264.

35. Hasegawa G, Hirano M, Ishihara Y. Differential gene expression associated with inflammation and blood pressure regulation induced by concentrated ambient particle exposure. *Inhal Toxicol*. 2011;23:897–905.

36. Roy A, Gong J, Thomas DC, et al. The cardiopulmonary effects of ambient air pollution and mechanistic pathways: a comparative hierarchical pathway analysis. *PLoS One*. 2014;9:e114913.

37. Harkema JR, Keeler G, Wagner J, et al. Effects of concentrated ambient particles on normal and hypersecretory airways in rats. *Res Rep Health Eff Inst*. 2004;120:1–68; discussion 69–79.

38. Urich B, Silverman F, Corey P, et al. Acute blood pressure responses in healthy adults during controlled air pollution exposures. *Environ Health Perspect*. 2005;113:1052–1055.

39. Nightingale JA, Maggs R, Cullinan P, et al. Airway inflammation after controlled exposure to diesel exhaust particulates. *Am J Respir Crit Care Med*. 2000;162:161–166.

40. Gong H Jr, Wong R, Sarma RJ, et al. Cardiovascular effects of ozone exposure in human volunteers. *Am J Respir Crit Care Med*. 1998;158:538–546.

41. Zhao A, Chen R, Wang C, et al. Associations between size-fractionated particulate matter and blood pressure in a panel of type II diabetes mellitus patients. *Environ Int*. 2015;80:19–25.

42. Jacobs L, Emmerenchts J, Mathieu C, et al. Air pollution related prothrombotic changes in persons with diabetes. *Environ Health Perspect*. 2010;118:191–196.

43. Park SK, Wang W, Ambient Air Pollution and Type 2 Diabetes: a systematic review of epidemiologic research. *Curr Environ Health Rep*. 2014;1:275–286.

44. Rudež G, Jansen NAH, Kiline E, et al. Effects of Ambient Air Pollution on Hemostasis and Inflammation. *Environ Health Perspect*. 2009;117:995–1001.

45. Nemmar A, Zia S, Subramaniyan D, et al. Exacerbation of thrombotic events by diesel exhaust particles in mouse model of hypertension. *Toxicology*. 2011;285:39–45.

46. Yitshak-Sade M, Kloo6, I, Liberty IF, et al. Air pollution and serum glucose levels: a population-based study. *Medicine*. 2015;94:e1093.

47. Guo Y, Li S, Tawatsupa B, et al. The association between air pollution and mortality in Thailand. *Sci Rep*. 2014;4:5509.

48. Yi O, Hong YC, Kim H. Seasonal effect of PM(10) concentrations on mortality and morbidity in Seoul, Korea: a temperature-matched case-crossover analysis. *Environ Res*. 2010;110:89–95.

49. Neuberger M, Moshammer H, Rabczenko D. Acute and subacute effects of urban air pollution on cardiopulmonary emergencies and mortality: time series studies in Austrian cities. *Int J Environ Res Public Health*. 2013;10:4728–4751.

50. Zhou M, He G, Liu Y, et al. The associations between ambient air pollution and adult respiratory mortality in 32 major Chinese cities. *Environ Res*. 2015;137:278–286.

51. Gosepath J, Schaefer D, Brommer C, et al. Subacute effects of ozone exposure on cultivated human respiratory mucosa. *Am J Rhinol*. 2000;14:411–418.

52. Ng CF, Ueda K, Niita H, et al. Seasonal variation in the acute effects of ozone on premature mortality among elderly Japanese. *Environ Monit Assess*. 2013;185:8767–8776.

53. Wang G, Jiang R, Zhao Z, et al. Effects of ozone and fine particulate matter (PM2.5) on rat system inflammation and cardiac function. *Toxicol Lett*. 2013;217:23–33.

54. Liu Y, Chen X, Huang S, et al. Association between air pollutants and cardiovascular disease mortality in Wuhan, China. *Int J Environ Res Public Health*. 2015;12:3506–3516.

55. Marti-Soler H, Gonseth S, Gubelmann C, et al. Seasonal variation of overall and cardiovascular mortality: a study in 19 countries from different geographic locations. *PLoS One*. 2014;9:e113500.

56. Williams ML, Atkinson RW, Anderson HR, et al. Associations between daily mortality in London and combined oxidant capacity, ozone and nitrogen dioxide. *Air Quality, Atmosphere Health*. 2014;7:407–414.