Study of physico-chemical and biochemical parameters during rotary drum composting of water hyacinth

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Abstract

Background Water hyacinth (Eichhorniacrassipes) is one of the most uncompromising weeds in the whole world. Its adverse effects due to fast growth rate are main physical interference with fishing and navigation. Water hyacinth also causes eutrophication due to the large release of organic nutrients after its degradation, consequentially deterioration of water quality and also adversely affecting aquatic flora and fauna. Therefore, composting is one of the best methods for control and utilization of water hyacinth. Water hyacinth being the plant material is rich in cellulose, hemicellulose, and lignin content which hinders the rate of degradation during composting. The raw materials including water hyacinth along with sawdust and cattle manure in five different proportions trial 1 (10:0:0), trial 2 (8:1:1), trial 3 (7:2:1), trial 4 (6:3:1), and trial 5 (5:4:1) were composted using rotary drum composter.

Results Final product of water hyacinth composting was flourishing of nutrients such as nitrogen, phosphorus, sodium, potassium, calcium, and magnesium. The lignin reduction in all the five trials was obtained between 10 and 40 %. The reduction in cellulose was observed ranging from 4 to 55 % in different trials. Similar as cellulose and lignin, hemicellulose was also reduced about 11–46 % in all five trials during the process.

Conclusion The maximum reduction inorganic matter, lignin, cellulose and hemicellulose was observed in trial 4; whereas, the nutrient contents (nitrogen, phosphorus, Na, K, Ca, and Mg) were increased significantly during the process. On analyzing the FTIR results, trial 4 showed that aliphatic and polysaccharides have easily degraded and aromatic compounds have increased with composting time in trial 4.

Keywords Water hyacinth · Rotary drum composter · Lignin · Cellulose · Hemicellulose · Reducing sugars

Introduction

Water hyacinth (Eichhorniacrassipes) or pani-meteka in Assam, India is found abundantly in almost every water body of the Assam state. It is an inherent of Brazil and other central South American countries, but now it is widespread in lakes, slowly moving rivers and swamps in most countries of the world lying between 40°N and 40°S includes India, South Africa, and the USA (Malik 2007). Water hyacinth has seen as one of the most productive plants in the world and it has many advantages for human beings (Abbasi and Ramasamy 1999). It has been used for biogas production, dye removal, as a source of biofuel, bioelectricity generation, animal feed, agriculture, and household articles manufacturing (Malik 2007). Its function as a food source for aquatic bio-phage, water current controls, purifies turbid water through sedimentation and sorption, and reduces pollutants through absorption of minerals (Baruah 1984). The neutral pH, high concentration of macronutrients in water, optimum temperatures (28–30 °C) and high intensity of light are favorable conditions for flourishing water hyacinth (Center et al. 2002). Due to its fast growth, it causes several problems in the affected water bodies such as, physical interference with
fishing, navigation, irrigation systems, hydroelectric power schemes and increased sedimentation and blockage in canals, rivers can give rise to risk of flooding (Malik 2007). Attempts to control this weed have caused high expenses and labour requirements, which only aids in temporary removal of this weed. If it is not controlled and utilized properly, aquatic ecosystem will be affected adversely. But, the water hyacinth possesses a great ability if seen as raw material for industries or if assimilated into agricultural practices (Gunnarsson and Petersen 2007). Therefore, composting can be seen as one of the sustainable option for utilization of water hyacinth into agricultural field.

Water hyacinth being a plant biomass is primarily composed of three constituents such as lignin, cellulose, and hemicellulose. This plant is fibrous in nature, and contains a high energy and protein content, due to which it can be utilized for numerous practical applications (Malik 2007). The amount of hemicellulose and cellulose is moderately high in water hyacinth plant but these substances are hardly available for microorganisms due to strong association with the lignin present in the plant cell. Since, the biodegradation of lignin is very difficult in the nature (Patel et al. 1993). The lignocellulose is the major organic compound that limits the rate of composting of agricultural and forestry waste, among which the lignin acts as a major contributor of speed limiting compound (Patel et al. 1993).

Lignin degradation is predominantly an aerobic process, and in an anaerobic condition, lignin can persevere for longer time (Van 1994). Lignin being the most recalcitrant component of the plant cell wall, its higher proportion reduces bioavailability of the substrate (Haug 1993). The most abundant constituent of plant is the cellulose and hemicelluloses, which can be easily degraded to a large extent during composting period, but a portion of it is associated with lignin molecule; maximum degradation could not be achieved. Therefore, to achieve the quality of water hyacinth compost, it is required to study the degradation of lignin, cellulose and hemicelluloses during the composting process without inoculating it with external lignin degrader. Rotary drum composting of water hyacinth being the method which provides high-rate composting of as a result of proper mixing of compost mixture and higher thermophilic temperature; therefore it has been adopted to achieve high degradation rate (Singh and Kalamdhad 2013a, b, 2014). Hence, five different trials were performed with different proportions of water hyacinth, sawdust and cattle manure and the best proportion of these materials was evaluated for achieving maximum reduction of lignin, cellulose and hemicellulose under ideal conditions.

Materials and methods

The compost materials

Water hyacinth was collected from the industrial area situated in Amingaon, North Guwahati, Assam, India. Cattle manure and sawdust were obtained from nearby dairy farm and saw mill, respectively. The water hyacinth was collected and then shredded into size of 1 cm approximately for uniform substrate distribution. Cattle manure and sawdust were properly mixed together in the desirable ratio and finally mixed with shredded water hyacinth. Five trials (Trials 1, 2, 3, 4, and 5) were performed during rotary drum composting of water hyacinth in combination with cattle manure and sawdust. Sawdust was used as a bulking agent. Mixtures of these materials were composted in rotary drum composter for 20 days. The drum was rotated after every 24 h for proper aeration and homogenized mixing of materials for degradation. The compost was prepared with different proportions of waste composition as described in Table 1.

Rotary drum composter design

A rotary drum composter of 550 L capacity was used as adopted by Kalamhdad et al. (2009a, b). The main unit of the composter, i.e., the drum is of 0.92 m in length and 0.9 m in diameter, made up of a 4 mm thick metal sheet.

| Table 1 Waste composition and initial characterization of raw materials |
| --- |
| Trials | Raw materials |
| Water Hyacinth (kg) | Cattle manure (kg) | Saw dust (kg) |
| Trial 1 (10:0:0) | 150 | 0 | 0 |
| Trial 2 (8:1:1) | 120 | 15 | 15 |
| Trial 3 (7:2:1) | 105 | 30 | 15 |
| Trial 4 (6:3:1) | 90 | 45 | 15 |
| Trial 5 (5:4:1) | 75 | 60 | 15 |
| pH | 5.79 | 6.65 | 6.15 |
| Moisture content (%) | 85.93 | 80.92 | 10.01 |
| Organic matter (%) | 72.64 | 97.87 | 72.05 |
| Acid insoluble lignin (%) | 9.65 ± 0.23 | 20 ± 2.12 | 41.35 ± 1.17 |
| Acid soluble lignin (%) | 1.80 ± 0.10 | 1.27 ± 0.02 | 0.86 ± 0.02 |
| Cellulose (%) | 15.78 ± 0.51 | 49.87 ± 2.02 | 37.89 ± 2.04 |
| Hemicellulose (%) | 27 | 11.8 | 4.3 |
| Reducing sugars (%) | 0.65 ± 0.01 | 1.386 ± 0.38 | 1.942 ± 0.12 |
The drum was rotated every 24 h for proper agitation and aeration of raw materials inside the composter. The samples were collected through grab sampling from different locations, mainly from the mid span and end terminals of the rotary drum composter with a compost sampler to minimize disturbance of the adjacent materials. All the grab samples were thoroughly mixed together to make a homogenized sample. Triplicates homogenized samples were collected on 0, 4, 8, 12, 16, and 20th day after manual turning.

The analytical methods

Temperature was measured daily using a digital thermometer before turning the compost material. Triplicate samples were collected and dried at 105 °C in hot air oven for 24 h and moisture content was calculated, dried samples were ground to pass to 0.2 mm sieves and stored for further analysis. Later, the samples were analyzed for following parameters: pH (1:10, w/v waste:water extract) and volatile solids (VS) by combustion at 550 °C for 2 h. Total nitrogen was measured using the Kjeldahl method, NH₄–N and NO₃–N using KCl extraction (Tiquia and Tam 2000); available total phosphorus (acid digest) was analyzed using the stannous chloride method (APHA 2005). The Flame photometer (Systronic 128) was used for analysis of Na, K and Ca concentration and atomic absorption spectrometer (AAS) (Varian Spectra 55B) was used for analysis of Mg after the digestion of 0.2 g sample with 10 mL of H₂SO₄ and HClO₄ (5:1) mixture in block digestion system (Pelican Equipments Chennai, India) Pelican for 2 h at 300 °C.

Lignin was determined in 0.3 g (dry weight) portions of each sample using National renewable energy laboratory (NREL) procedure (Templeton and Ehrman 1995; Ehrman 1996). Cellulose was obtained using method adopted by Updegraff (1969) and hemicellulose was determined from the difference between neutral detergent fiber (NDF) and acid detergent fiber (ADF) using method provided by Goering and Van (1975). Reducing sugars were obtained using DNS method Miller (1959).

**Results and discussions**

**Physico-chemical analysis**

Temperature profiles for all the five trials are shown in Fig. 1. The rise in temperature was caused mainly by the metabolic heat generated during degradation of organic matter. In trial 1 (control), the maximum temperature recorded about 36.2 °C on day 6, which was the least peak achieved as compared to other trials due to only water hyacinth was used for composting (without addition of cattle manure and sawdust). The maximum temperature recorded by trials 2, 3, 4, and 5 was 50.1, 51.9, 59.7, and 53.6 °C on sixth day, respectively. The highest peak of temperature (59.7 °C) was observed in trial 4 with waste composition of water hyacinth, cattle manure and sawdust (6:3:1) due to optimum proportion of raw materials. All the composting trials accomplished ambient temperature at the end of 20th day indicating the maturity of compost. It also indicated that the microbial activity has reduced due to decrease in the amount of degradable organic matter. Similar results have been obtained by Dhal et al. (2012) where maximum temperature was recorded as 57.3 °C. Singh and Kalamdhad (2012) reported the maximum temperature of 56 °C during water hyacinth composting.

Moisture loss during the composting process can be viewed as an index of decomposition rate, because the heat generated is directly proportional to the evaporation rate of moisture in compost. However, optimum moisture content is necessary for the growth and survival of microorganisms. Higher moisture loss was observed in trial 5 (27.42 %) followed by trial 1 (23.53 %), 4 (19.69 %), 3 (19.17 %), 2 (8.91 %), respectively, at the end of composting period of 20th day (Fig. 2a). Leachate formation was not observed during composting period.

The pH of water hyacinth was observed between 6 and 8. Therefore, the composition of raw materials (cattle manure, sawdust, and water hyacinth) was taken such that the resultant pH lies in between 6 and 7.5, which is the most favorable condition for the growth of microorganisms. Nakasaki et al. (1993) tested the pH dependency of organisms active in the composting process and found the pH range of 7–8 to be optimum, whereas according to Bharadwaj (1995), the optimum pH for most microorganisms is between 6.5 and 7.5. Figure 2b shows the variation of pH during composting of different trials. As per the results obtained, the initial pH for trials 1, 2, 3, 4, and 5 was 6.87, 5.37, 6.62, 6.49, and 6.63, which further increased to 6.96, 7.09, 6.77, 6.80, and 7.37, respectively, at the end of composting period. Also, a slight decrease in pH was observed in all the trials between day 8 and day 12. But, for
Lignin degradation, it was observed that pH 4.5 was optimum. The degradation suppresses below pH 3.5 and above 5.5 (Kirk et al. 1978). The pH was not observed less than 6.2 in all the trials except the trial 1 (control). Hence, pH has not contributed for any lignin degradation. The slight decrease in pH during the initial period of composting was expected due to the formation of acid as a result of metabolism of readily available carbohydrates. Overall, the acceptable pH ranges were within levels tolerable to microorganisms. Neutral pH range was obtained at the end of composting in all the trials.

Figure 2c shows the degradation of organic matter during composting process. The maximum loss of organic matter was achieved in trial 4 (32.58 %). Most of the organic matter degraded during composting and consequently more oxygen consumed in trial 4. The percentage loss of organic matter in trials 1, 2, 3, and 5 was found as 7.96, 10.62, 27.95, and 24.52 %, respectively. At the end of composting period the amount of organic matter in the compost was determined to be 74, 74, 61.6, 43.2 and 62.5 in trials 1, 2, 3, 4 and 5, respectively. Organic matter is mostly degraded during thermophilic phase of composting where oxygen is consumed at higher rate by the microorganisms. Hence, it can be concluded that for organic matter degradation, temperature and moisture play a very significant role.

Figure 3a shows that total nitrogen (TN) was increased from 0.91 to 1.19 %, 0.63 to 0.98 %, 0.58 to 0.91 %, 0.68 to 1.05 %, and 0.91 to 1.54 % in trials 1, 2, 3, 4, and 5, respectively, within composting time. Highest TN was increased in trial 5 (69.23 %) followed by trial 3 (56.99 %), trial 2 (55.56 %) and trial 4 (54.41 %) and trial 1 (30.77 %) during the composting process. TN was increased in rotary drum composting due to the net loss of dry mass in terms of CO₂ as well as the water loss by evaporation as a result of heat evolution during oxidization of organic matter (Kalamdhad et al. 2009a, b). The concentration of NH₄–N was decreased from 97.8 to 77.1 mg/kg in trial 1, from 155.6 to 68.5 mg/kg in trial 2, from 201.8 to 65.7 mg/kg in trial 3, from 148.5 to 22.7 mg/kg in trial 4, and 69.7 to 43.4 mg/kg in trial 5 during the composting process (Fig. 3b). Higher reduction of NH₄–N was observed in trial 4 (84.7 %) followed by trial 3 (67.4 %), trial 2 (56.0 %), trial 5 (37.7 %), and trial 1 (21.1 %) during the composting process (Fig. 3). The absence or decrease in NH₄–N is an indicator of both high-quality composting process. High pH, mixing and increased aeration have been exposed to enhance ammonia loss during maturation (Kalamdhad et al. 2009a, b). On analyzing the results by ANOVA, significant differences in nutrients were observed between the trials (P < 0.05). The total phosphorus was increased from 2.4 to 3.0, 2.4 to 4.0, 2.8 to 4.1, 2.3 to 4.6, and 2.6 to 3.7 g/kg in trials 1, 2, 3, 4 and 5, respectively (Fig. 3c). The available phosphorus was also increased from 0.7 to 0.9, 1.2 to 2.4, 1.5 to 2.1, 1.4 to 2.0, and 1.2 to 1.7 g/kg in trials 1, 2, 3, 4, and 5, respectively, during the composting process (Fig. 3d). Increase in total and available phosphorus concentration indicating the higher microbial activities during the composting process caused more mineralization; during mineralization process, net loss of compost mass was reduced (Kalamdhad et al. 2009a, b). The concentration of nutrients such as Na, K, Ca, and Mg was increased in the range of 12–120, 9–72, 4–133, and 6–72 %, respectively, in all the trials during the process; it might be due to the net loss of dry mass (Singh and Kalamdhad 2013a).

Biochemical analysis

Lignin (acid soluble and insoluble)

Figure 4a shows the degradation of acid insoluble lignin in the compost environment. Lignin has a very complicated structure, which is highly resistant to microbial degradation (Hatakka 2001). Its degradation mainly occurs during
thermophilic phase of composting (Tuomela et al. 2000) and an optimum temperature of about 40–50 °C is required for the growth of thermophilic fungi and actinomycetes for a longer duration. Therefore, rotary drum composting method was adopted to achieve a high rate of degradation in amount of lignin present. Acid insoluble lignin was increased about 13.95% in trial 1 (control), whereas, it was reduced about 9.09, 16.16, 39.72, and 10.29% in trials 2, 3, 4 and 5, respectively, during the composting process. During initial characterisation of raw materials, the percentage of acid insoluble lignin was found about 41.34 and 20.0% in sawdust and cattle manure, respectively. These raw materials contributed to the increase in amount of acid insoluble lignin present in the compost. The maximum biodegradation of acid insoluble lignin was observed in trial 4 (39.72%) as compared to other trials. It can be attributed as; proper proportion of initial waste mixture enhanced the growth and activity of microorganisms. Since, lignin is difficult to degrade, the percentage of un-degraded acid insoluble lignin was found to be 28.67, 35.67, 46.67, 28.33, and 36.33% in trials 1, 2, 3, 4, and 5, respectively, during the composting process. Regardless of adopting a high rate of composting method, the maximum reduction of 39.72% in acid insoluble lignin could be achieved during water hyacinth composting. Proper aeration by manual turning consequently achieving maximum thermophilic temperature was also contributing factors for the degradation of acid insoluble lignin. An optimum temperature of about 40–50 °C is required for thermophilic micro-fungi and actinomycetes to degrade lignin but, above 60 °C, these microorganisms cannot grow and lignin degradation is slowed down (Tuomela et al. 2000). Therefore, complete degradation of lignin is not accomplished in any of the trials within 20 days. Feng et al. (2011) reported degradation of 11.74% of lignin during composting of lignocellulose material. Garcia et al. (2003) observed 60% decreases in lignin amount during composting of olive mill waste. Figure 4b shows the degradation of acid soluble lignin. The higher reduction of acid soluble lignin was observed in trial 3 (35.83%) followed by trial 4 (25.89%), trial 2 (16.13%), trial 5 (7.92%), and trial 1 (3.32%) during the composting process. The percentage of acid soluble lignin were observed about 2.71, 1.30, 0.77, 0.83, and 1.86% in trials 1, 2, 3, 4, and 5, respectively, in the final compost, which was very less as compared to insoluble lignin.

**Cellulose**

Cellulose is the major element of plant cell walls which provides them rigidity (Beguin and Albert 1994). Cellulose was found high in cattle manure (49.87%), whereas in water hyacinth and sawdust, the percentage of cellulose was 15.78 and 37.89%, respectively. Combination of these
three raw materials might have resulted into higher amount of cellulose in all the trials. Since, the cellulose is easily degradable organic matter and its degradation starts the moment raw materials are piled up (Cooperband 2000). Therefore, the initial amount of cellulose present in trials 1, 2, 3, 4, and 5 was 33.25, 36.69, 41.89, 32.92, and 31.6 %, respectively. Figure 4c shows the degradation of cellulose during composting of water hyacinth. The highest degradation for cellulose was observed in trial 4 (55.07 %) due to proper combination of waste materials which resulted in better biological activity. The slow degradation of cellulose could be explained as the lignin attached strongly with cellulose (Kuhad et al. 1997). The degradation of cellulose in trials 1, 2, 3, 4 and 5 was obtained as 26.25, 9.02, 3.65, 55.07 and 23.16 %, respectively. At the end of composting period, the amount of un-degraded cellulose left in compost was found to be 24.52, 33.38, 40.36, 14.79 and 24.28 %, respectively, in trials 1, 2, 3, 4 and 5. Mesophilic bacteria, actinomycetes, fungi, and protozoa are the microorganisms which grow at temperature between 10 and 45 °C i.e., during mesophilic phase to biodegrade raw materials (Cooperband 2000) and break down easily degradable components such as sugars and amino acids (Hellmann et al. 1997). Parveen and Padmaja (2010) reported cellulose degradation of 31.55 % in uninoculated and 20.56 % in fungal consortium inoculated during composting of water hyacinth and municipal solid waste. Baddi et al. (2004) reported 58 % reduction in cellulose in 12 months during olive mill composting.

**Hemicellulose**

Hemicellulose is a complex carbohydrate polymer and makes up 25–30 % of total wood dry weight. It primarily consists of five sugars D-xylose, D-mannose, D-galactose, D-glucose, and L-arabinose. It also has 4-O-methyl-glucuronic, D-galacturonic, and D-glucuronic acids. Sugars are linked together by β-1, 4- and occasionally β-1, 3-glycosidic bonds (Perez et al. 2002). In contrast to cellulose, they are easily hydrolyzable polymers. They do not form aggregates, even when they are co-crystallized with cellulose chains. The maximum amount of 27 % of
hemicellulose was present in water hyacinth. The highest reduction of 45.16 % in hemicellulose was obtained in trial 4. For trials 1, 2, 3, and 5, the reduction was observed as 39.66, 32.03, 11.65, and 43.96 %, respectively (Fig. 4d). Baddi et al. (2004) reported 76 % reduction in hemicellulose in 12 months during olive mill composting. Zeng et al. (2010) reported maximum degradation of hemicellulose as compared to lignin and cellulose during composting. In all the five trials, it was found that the amount of hemicellulose present in the compost at the end of 20 days was less than 10 % indicating very rapid degradation of hemicellulose under microbial activity.

Reducing sugars

Sugars with reducing property arising out of the presence of a potential aldehyde or keto group are called reducing sugars. Some of the reducing sugars are glucose, galactose, lactose, and maltose. In trial 1, the amount of reducing sugars present initially was 21.61 %. Maximum reduction of 98.75 % in reducing sugar was observed in trial 1 (control) in which no bulking agents were present. For trials 2, 3, 4, and 5, the degradation of reducing sugar is 62.38, 67.3, 77.1, and 47.3 %, respectively (Fig. 4e). The reduction in reducing sugars may be attributed due to microbial assimilation as the composting proceeded. Parveen and Padmaja (2010) reported that reduction in reducing sugar was 0.67 mg/g in water hyacinth compost.

$E_4/E_6$ ratio

Formation of humic substance is important parameter of the mature compost; therefore, to study the changes in humification parameter $E_4/E_6$ ratio is widely adopted. It provides the quality of humic acid and aromatization level of compost. If the E4/E6 ratio is below five then the sample is characterized as humic acid, whereas if the ratio is above five then the sample is characterized as fulvic acid (Zorbas et al. 2008). In Fig. 4f, it was observed that E4/E6 ratio has increased in trials 1, 3, 4 and 5. The magnitude of E4/E6 is associated with the degree of condensation of the aromatic carbon network; also it is independent of concentrations of humic materials (Zorbas et al. 2008). A low E4/E6 ratio indicates relatively high degree of condensation of aromatic humic constituents, whereas a high E4/E6 ratio shows a low degree of aromatic condensation and implies the presence of relatively large portions of aliphatic structures (Rivero et al. 2004). Low degree of aromatic carbon might have resulted due to short period of humification during composting process. The value of E4/E6 ratio for humic acids usually ranges between 2 and 5 and the results have been obtained within this range. Humic acids derived from compost usually have low degree of aromatic carbon and low-molecular weight (Velasco et al., 2004). Since, this E4/E6 ratio should decrease with time; the same can be achieved if time allowed for compost maturity is increased. Rivero et al. (2004) observed a decrease in E4/E6 ratio on 3-year-old compost. Huang et al. (2006) achieved a decrease in E4/E6 ratio from 3.53 to 2.50 in 63 days. Wei et al. (2007) also observed a decrease in value from 8.1 to 6.1 at the end of 63 days of MSW compost.

FTIR spectroscopy

Figure 5 shows IR spectra of trials 1, 2, 3, 4 and 5 during composting process. IR spectra were observed for all the five trials with variations in peak intensity of the band. A broad band at 3,400–3,300 cm$^{-1}$ confirmed the hydrogen variations of alcoholic (OH), phenolic or carboxylic groups (COOH). Absorption peaks were observed at 2,930–2,920 cm$^{-1}$ (C–H stretching vibrations of fatty acids, waxes and various aliphatic components). Peaks at 2,850 cm$^{-1}$ show C–H stretching vibrations in the presence of aliphatic methylene. Peaks at 1,630 cm$^{-1}$ characterize aromatic C=C vibrations and alkenes. Peaks at 1,510–1,520 cm$^{-1}$ are assigned to aromatic skeletal of lignin and other lignocellulose components. Peaks at 1,430–1,420 cm$^{-1}$ respond to signals generated by –CH=, –CH$_2$ and CH$_3$ radicals. Carbone are also known for absorption at 1,420 cm$^{-1}$. Peaks at 1,380 cm$^{-1}$ confirm the presence of mineral constituents (N–O stretch). Ammonium bicarbonate is formed by reaction between ammonia and CO$_2$ released during composting process. 1,320 cm$^{-1}$ absorbs C–N stretch with components such as primary aromatic and secondary amines. Peaks from 1,032 to 1,040 cm$^{-1}$ absorb Si–O–Si silicates, aromatic ethers and polysaccharides (C–H stretch). The interpretation of spectra is based on work published by Baddi et al. (2004) and Singh et al. (2011).

Changes were observed only in the band intensity during the composting process, which suggests that compost composition was same throughout the composting period. A very slight reduction in aliphatic C peak at 2,930 cm$^{-1}$ was observed, shoulder peak at around 2,850 cm$^{-1}$ smoothened and disappeared by the end of the composting process in most of the trials. This can be explained by the microbes using aliphatic and peptide structures and carbohydrates such as polysaccharides, cellulose and hemicellulose to meet their energy needs. Quantitative evaluation of compost sample with different proportionates was done in terms of relative intensity ratios of selected absorbance band. Decrease in 2,930/1,030 (aliphatic Carbon/polysaccharide) was observed in trials 3 and 4. The ratio of intensities decreased from 0.87 to 0.79 and 0.68 to 0.63, respectively. The ratio of intensities 1,630/1,030 (aromatic carbon/polysaccharide) decreased from 0.90 to 0.85, 0.90 to 0.85, 0.75 to 0.74, and 0.90 to 0.88 in trials 1, 3, 4 and 5, respectively. Similarly, decrease in intensity
Fig. 5 IR spectroscopy of compost samples: a trial 1, b trial 2, c trial 3, d trial 4, and e trial 5
ratio 2.930/1.630 (aliphatic carbon/aromatic carbon) was observed as 0.82 to 0.78, 0.96 to 0.93, and 0.91 to 0.84, in trials 1, 3 and 4, respectively. This shows that with increase in composting period, the compost was enriched in aromatic C=C as compared with aliphatic carbon. Kalamdhad et al. (2009a, b) observed the similar trend in the ratio of intensities.

**Conclusion**

Rotary drum composter was extremely competent for making water hyacinth compost due to providing proper mixing, agitation and aeration for composting microorganisms. The maximum reduction in lignin, cellulose and hemicellulose was observed in trial 4 (6:3:1) in comparison to trials 1, 2, 3 and 5 during rotary drum composting of water hyacinth. In addition, maximum temperature and organic matter reduction was also observed in trial 4. The nutrients contents (nitrogen, phosphorus, Na, K, Ca and Mg) were increased significantly during the process. It was observed that hemicellulose was degraded at a faster rate as compared to cellulose and lignin, since it is easily hydrolysable substance under mesophilic stage of composting process. From the results of $E_d/E_o$ ratio, it was concluded that the time required for compost maturity can be increased; since, humification increases with maturity of compost. On analyzing the FTIR results, trial 4 showed that aliphatic and polysaccharides have easily degraded and aromatic compounds have increased with time in trial 4. Hence, it can be concluded that the trial 4 (6:3:1) could be beneficial to enhance the agronomic value of soil as compared to other trials. Furthermore, to achieve higher degradation of cellulose and lignin, combination of rotary drum with microbial inoculation can be proposed.

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