Density-Dependent Demographic Variation Determines Extinction Rate of Experimental Populations

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Understanding population extinctions is a chief goal of ecological theory. While stochastic theories of population growth are commonly used to forecast extinction, models used for prediction have not been adequately tested with experimental data. In a previously published experiment, variation in available food was experimentally manipulated in 281 laboratory populations of *Daphnia magna* to test hypothesized effects of environmental variation on population persistence. Here, half of those data were used to select and fit a stochastic model of population growth to predict extinctions of populations in the other half. When density-dependent demographic stochasticity was detected and incorporated in simple stochastic models, rates of population extinction were accurately predicted or only slightly biased. However, when density-dependent demographic stochasticity was not accounted for, as is usual when forecasting extinction of threatened and endangered species, predicted extinction rates were severely biased. Thus, an experimental demonstration shows that reliable estimates of extinction risk may be obtained for populations in variable environments if high-quality data are available for model selection and if density-dependent demographic stochasticity is accounted for. These results suggest that further consideration of density-dependent demographic stochasticity is required if predicted extinction rates are to be relied upon for conservation planning.

Citation: Drake JM (2005) Density-dependent demographic variation determines extinction rate of experimental populations. PLoS Biol 3(7): e222.

Introduction

As changing global climate and anthropogenic modification of the biosphere increasingly threaten global biodiversity [1,2], accurately predicting extinctions takes on new urgency [3–5]. Although stochastic models of population growth and decline are commonly used to forecast extinction, quantitative model predictions for population dynamics in randomly fluctuating (i.e., natural) environments have only been poorly supported by experiments [4,6–7]. In contrast, validation with observational field data [5,8–10] is beset by numerous difficulties, including low precision [11], low replication [12], small sample size [12,13], poor data quality [12], and lack of independence [12,14]. Without empirical validation, viability analyses for threatened and endangered species are subject to doubt, and identifying risk factors for population extinction is difficult. The results reported here support the further use of stochastic models for predicting extinction, but only where detailed information about variation in individual fitness is available for estimating the structure and magnitude of demographic stochasticity.

Two components contribute to random variation in per capita population growth rates: demographic stochasticity and environmental variability [15–18]. Here, demographic stochasticity refers to variation in individual fitness, sampling effects in finite populations, and chance events that affect individuals independently [18–20]. Variation in environmental conditions can cause expected individual fitness to fluctuate, resulting in environmental stochasticity. Defining \( \Delta N = N_{t+1} - N_t \), a model for per capita population growth rate with demographic and environmental stochasticity is

\[
\frac{\Delta N}{N} = [\lambda(N, \psi) - 1] + W(\sigma_e(N) + \sigma_d(N)/N)
\]

where \( \lambda \) is the (possibly density-dependent) expected multiplicative per capita population growth rate depending on population size \( N \) in the previous time step and parameters \( \psi, W \) is a normal unit variate, and \( \sigma_e(N) \) and \( \sigma_d(N) \) are models for the (possibly density-dependent) effects of environmental and demographic stochasticity, respectively [18]. This is a general model, which can approximate or has as its special cases many common stochastic population growth models, including linear and nonlinear stochastic differential equations [21,22], birth-death processes [17,23,24], and discrete and continuous-time branching process models [25], although it is unclear if these approximations reliably represent the dynamics of small populations [26].

Recent research suggests that demographic stochasticity might have a density-dependent component in the sense that the variance contributed to the population's growth rate by demographic stochasticity, represented by \( \sigma_d \) in Equation 1, is a function of population size [18,19,27]. This phenomenon is independent of the well-known scaling of demographic stochasticity with population size, represented by the ratio \( \sigma_d(N)/N \) in the model [15,16,18,25]. While theory predicts that...
density-dependent demographic variation will occur in all populations in which vital rates are density-dependent [19], and will therefore be ubiquitous in both natural and experimental populations, the severity of density dependence in demographic variation has rarely been investigated, and thus rarely measured [18,27]. Extinction forecasts that ignore density-dependent demographic stochasticity in populations in which it is present will be biased, possibly severely.

To test for effects of environmental variation on persistence, 281 laboratory populations of water fleas (*Daphnia magna*) were maintained under experimentally controlled conditions for 104 d using three levels of variation in food availability as a source of environmental stochasticity [28]. Here, those data are used to select and test stochastic models of population dynamics. To ensure independence, the data were divided into subsets for model fitting and model testing.

Then, as an exploratory analysis, deviations were plotted from expected population growth against population size for populations in the low-variability treatment to suggest how demographic variation might be structured. On the basis of this analysis, a set of models was developed for population growth in fluctuating and constant environments to test for effects of environmental variation. Finally, population growth trajectories with (i) density-dependent demographic stochasticity, and (ii) density-independent demographic stochasticity were simulated to obtain model predicted extinction rates. Simulated extinctions for the model with density-dependent demographic stochasticity were consistent with observed rates of extinction in the test dataset or were only slightly biased, while the model with density-independent demographic stochasticity considerably underestimated observed extinctions. These results suggest that biased estimates of extinction risk resulting from ignoring density-dependent demographic stochasticity may lead to unwarranted optimism concerning the eventual fates of endangered populations.

**Results**

**Exploratory Graphical Analysis**

After rescaling to isolate density dependence in \( \sigma_d \) (see Materials and Methods), the scatter plot of observed deviations from expected population growth rate suggested that variation from demographic stochasticity was dependent on population size in populations in the low-variability treatment (Figure 1). Nonparametric tests for correlation confirmed that the negative relationship between \( \log_e(1 - \text{deviation}) \) and population size was highly significant (Spearman rank-order correlation: \( \rho = -0.23, p < 0.0001 \); Kendall's \( \tau: \rho = -0.17, p < 0.0001 \); \( N = 342 \)). The average deviation was fit to an exponential model (see Materials and Methods) (Table 1).

**Estimation of Extinction Rates in Experimental Populations**

Extinction rates of populations in experimental treatments with low and medium levels of variation were not significantly different, although these were different from the extinction rate of populations subject to a high level of environmental variation (see Materials and Methods). Consequently, observations in the testing dataset from the low and medium variability treatments were pooled for remaining analyses, resulting in a more conservative test of model-predicted extinction rates. The maximum likelihood estimate of extinction rate for populations from the low- and medium-variability treatments in the model-testing dataset was 48.4% (95% confidence interval [CI], 38.0%–58.9%; Figure 2). The maximum likelihood estimate of extinction rate in the high-variability treatment was 70.2% (95% CI, 55.1%–82.7%; Figure 2).

**Model Selection**

Model-predicted extinction rates were obtained by fitting a simple Ricker model for population growth, \( \hat{E}(\lambda) = \lambda e^{-\lambda N} \), to data in the model-fitting dataset, although a 9-Ricker model was also considered (see Materials and Methods). Moderate to considerable support was obtained for models with only demographic stochasticity compared to models with both demographic and environmental stochasticity, according to Akaike's information criterion (AIC; Table 1). However, there was overwhelming support for a model with density-dependent demographic stochasticity compared to a model with density-independent demographic stochasticity for populations in the low- and medium-variability treatments (\( \Delta \text{AIC} = \ldots \)).

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**Figure 1. Demographic Stochasticity Is Strongly Density-Dependent in Experimental Populations of *D. magna***

Deviations from expected population size were rescaled by multiplying the observed deviation by initial population size for the interval to isolate density dependence in \( \sigma_d \) (see Materials and Methods). Rescaled deviations are strongly dependent on population size (\( p < 0.0001 \)). Because observed deviations overlap, obscuring the pattern, points have been jittered in the dimension of the x-axis by the addition of a small amount of random noise.

DOI: 10.1371/journal.pbio.0030222.g001

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**Table 1. AIC Scores for Ricker Models of Population Growth**

| Experimental Treatment          | Model I | Model II | Model III |
|---------------------------------|---------|----------|-----------|
| Low and medium variability (pooled) | 1026.8 (0) | 1028.8 (2.0) | 1063.8 (37.0) |
| High variability                | 518.6 (0) | 519.6 (1.0) | 522.8 (4.2) |

Models with density-dependent demographic stochasticity (model I), density-dependent demographic stochasticity and environmental stochasticity (model II), and density-independent demographic stochasticity (model III) were fitted to data from high-variability and pooled low- and medium-variability treatments. The difference between the model score and the score of the best model (\( \Delta \text{AIC} \)) is given in parentheses.

DOI: 10.1371/journal.pbio.0030222.t001
The accuracy of model predictions was improved by relaxing the usual assumption that demographic stochasticity is density-independent [21,25]. Since individual fecundity is commonly altered in response to population size or density [29,30], density-dependent demographic stochasticity resulting from demographic covariation is probably not exceptional in natural populations. Indeed, density-dependent demographic stochasticity will result from any density dependence in vital rates or interactions among individuals that affect demography [19]. Presently, this aspect of population biology is poorly understood (compare [18,27]). Additionally, the accuracy and precision of these predictions were facilitated by a relatively large, high-quality dataset \( (n_{\text{low}} = 342, n_{\text{med}} = 300, \text{and } n_{\text{high}} = 280) \). In general, field data will not be so abundant. Thus, an important goal for population biology is to develop methods for obtaining reliable predictions from sparse, low-quality datasets [31].

Planning for increasing threats to rare species from diverse sources, including climate change, resource extraction, habitat modification, and invasive species will require greater and more precise estimates of extinction risk than ever before. While the reliability of theoretical models for predicting extinction in natural ecosystems remains to be established, the results presented here show that accurate predictions of population extinction in variable environments are indeed possible.

**Materials and Methods**

**Experiment.** Experimental microcosms \( (n = 281) \) of *D. magna* were maintained on a food resource of *Selenastrum* sp. for 104 d. The daily food availability was experimentally varied at three levels (coefficient of variation = 0, 1, 2), which are referred to as “low,” “medium,” and “high,” respectively, while the long-run average volume of food over time was kept constant across all replicates in all treatments. Extinctions were tabulated and populations were counted daily, although populations with ten or more individuals were simply marked “abundant.” With the exception of these observations, sampling error is negligible. For more detailed methods, see [28]. Prior to this analysis, these data were filtered, retaining only observations from every seventh day for days 1 through 99, corresponding to the approximate generation time, resulting in observations of change in abundance over up to 14 intervals for each population. To achieve independence between data used for model fitting and data used for model testing, populations were assigned to separate datasets. Observations of ten or more individuals were excluded from the dataset for model selection and parameter estimation.

**Exploratory graphical analysis.** Because populations in the low-variability treatment were not exposed to any experimentally induced environmental variation, most variation in observed growth with density-dependent demographic stochasticity fit with independent data (triangles) accurately predicted extinction of populations in the low- and medium-variability treatments, but not the high-variability treatment. The addition of environmental stochasticity (square) improved the prediction, although the chance of obtaining the observed 33 extinctions (or more) out of 47 populations was only 3.3%. The standard model with constant demographic stochasticity (stars) fails to predict extinction in all treatments. DOI: 10.1371/journal.pbio.0030222.g002

**Figure 2.** Simple Models Accurately Predict Extinction in Experimental Populations

Estimates of the extinction rate in populations of *D. magna* reserved for model testing at three levels of environmental variation were obtained from the likelihood function of the binomial distribution (crosses, 95% CI). Because there was no difference between extinction rates in the low- and medium-variability treatments (see Materials and Methods), data were pooled to obtain a more precise estimate, resulting in a more conservative test (right of dashed line). Model-predicted estimates of extinction rate obtained from models of density-dependent population growth with density-dependent demographic stochasticity fit with independent data (triangles) accurately predicted extinction of populations in the low- and medium-variability treatments, but not the high-variability treatment. The addition of environmental stochasticity (square) improved the prediction, although the chance of obtaining the observed 33 extinctions (or more) out of 47 populations was only 3.3%. The standard model with constant demographic stochasticity (stars) fails to predict extinction in all treatments.

DOI: 10.1371/journal.pbio.0030222.g002

37.0) and considerable support for populations in the high-variability treatment \( (\Delta AIC = 4.2) \). These results confirm that density dependence in demographic variance strongly influenced realized population growth in experimental populations.

**Accuracy of Model-Predicted Extinction Rate**

The accuracy of model predictions can be assessed by comparing model-predicted extinction rates with the 95% CI for the estimated extinction rate of populations in the model-fitting dataset from the low-variability treatment. A scatter plot of \( ^{\prime}N_{t} \) versus \( N_{t} \) suggested that density dependence in expected population size is approximately linear on a logarithmic scale.
Estimating extinction rates in experimental populations. Because populations were independent, each population represents a Bernoulli trial for which the possible outcomes were extinction or persistence. Thus, the chance of extinction for a population in treatment level \( i \) is binomial with parameter \( p_i \). For \( n_i \) populations in the model-testing dataset, of which \( x_i \) were observed to go extinct, the maximum likelihood estimate of \( p_i \) and 95% CIs on the chance of extinction were determined from the likelihood function for the binomial distribution.

Initially, it was unclear if there was an effect of experimental treatment. Using regression on pooled observations from the entire experimental data set, Drake and Lodge [29] reported no significant effect of environmental variation on the chance of extinction (\( \phi = 0.0877 \)). Logistic regression was performed on the testing half of the dataset using treatment as a categorical covariate. The global null hypothesis was rejected by the likelihood ratio test (\( \phi = 0.0450 \)) but not by the \( \phi \) number of parameters being different than 0 (intercept, \( \phi_0 = 0.060; \phi_0 = 0.174; \phi_1 = 0.236 \)). Visual inspection of the data suggests that there might be no different between the low- and medium-variability treatments (Figure 2), but a difference between these and the high-variability treatment. Repeating logistic regression after pooling observations from the low- and medium-variability treatments unambiguously showed an effect of high variation on the chance of extinction (likelihood ratio test, \( \phi = 0.013 \); Wald's test, \( \phi = 0.015 \)). Therefore, to compare observed extinction rates with model-predicted extinction rates, observations from the low- and medium-variability treatments were pooled for both model fitting and model testing.

Model selection. The theoretical variance in \( \lambda \) is given by \( \sigma_\lambda^2/(\lambda N) \), while the mean can be given by any familiar population dynamical model such as Ricker, \( \theta \)-Ricker, \( \theta \)-logistic, or Gompertz growth. Using the exponential model for demographic variation and approximating the unknown distribution of \( \lambda \) with the first two moments, the negative log-likelihood function for the \( \theta \)-Ricker model of population growth \( \lambda(N) = \lambda_0 e^{-\theta N} \) is

\[
NLL(\lambda, \theta, \lambda_0, b_y(y)) = -\sum_{j=1}^{n_j} \left[ -\frac{2}{n_j} \log(\log(1 - e^{-\theta jN})) - \frac{\sum(y_j - (\lambda_0 e^{-\theta jN}))^2}{2(n_j^{\theta N}/N)} \right] \tag{2}
\]

where \( \lambda_0 \) and \( \theta \) can be interpreted as governing the intrinsic rate of increase and the severity of density dependence in the expected growth rate, respectively, and \( n_j \) is the number of observed intervals beginning at population size \( N \). The negative log-likelihood for this model with the addition of constant environmental stochasticity is

\[
NLL(\lambda, \theta, \lambda_0, b_y(y), \sigma_\lambda^2) = -\sum_{j=1}^{n_j} \left[ \frac{2}{n_j} \log(\log(1 - e^{-\theta jN})) - \frac{\sum(y_j - (\lambda_0 e^{-\theta jN}))^2}{2(n_j^{\theta N}/N)} \right] \tag{3}
\]

where \( \sigma_\lambda \) is a parameter representing the average level of variation from environmental stochasticity. Models were fit to data in the model-fitting dataset by minimizing the negative log-likelihood function using the Nelder-Mead simplex. Goodness of fit was quantified using AIC = 2NLL(\( \psi(y) + 2k \) where \( \psi \) is the vector of maximum likelihood estimates of all parameters, and \( k \) is the number of parameters estimated. The model with the lowest AIC score is the best fit after accounting for model complexity, while the relative support for a model with the lesser of two AIC scores differing by greater than 2 is suggestive of the second [32].

Overall, the inclusion of the parameter \( \theta \) (which allows for flexibility in the severity of density dependence in expected population growth rate) did little to fit the model and was fixed (\( \theta = 1 \)) for the remainder of the analysis, reducing the number of parameters to be estimated by one. AIC scores for remaining models are shown in Table 1. Interestingly, the estimates of \( \sigma_\lambda \) for data from all experimental treatments were not significantly different from 0, even for the treatment with a high level of experimentally induced variation.

Accuracy of model-predicted extinction rate. Predicted extinction rates were obtained by simulating 100,000 iterations of the Ricker model with density-dependent demographic stochasticity at maximum likelihood estimates of all parameters for populations in the pooled low- and medium-variability treatments and populations in the high-variability treatment separately using Euler's method [22]. Since predicted extinction rates for populations in the high-variability treatment were biased, and the model without environmental stochasticity was only weakly supported for these populations (Table 1), 100,000 iterations of the Ricker model were also simulated with density-dependent demographic stochasticity and environmental stochasticity for populations in the high-variability treatment.

The model for stochastic population growth considered here accounts for two factors commonly ignored when predicting the chance of population extinction: density-dependent changes in expected population size and density-dependent demographic stochasticity. Although the effect of density-dependence in \( E(X) \) on population persistence has been documented [25,33], the effects of density-dependent demographic stochasticity have not been as extensively studied [18,27]. Therefore, the chance of extinction that would have been predicted had the incorrect assumption been made that demographic stochasticity was density-independent was also determined. An estimate of density-independent demographic stochasticity for populations in the high-variability treatment was biased, and the model without environmental stochasticity was only weakly supported for these populations (Table 1), 100,000 iterations of the Ricker model were also simulated with density-dependent demographic stochasticity and environmental stochasticity for populations in the high-variability treatment.

As above, model predicted extinction rates were obtained by simulating 100,000 iterations of the extinction process at maximum likelihood estimates of all parameters.

Acknowledgments

The author thanks R. Schwartz, P. Baggenstos, and J. Frenness for assistance with this experiment; and J. Hellman, T. Coulson, M. Vellend, and two anonymous referees for comments on the manuscript. This research was supported by a Great Lakes Fishery Commission grant (to principal investigator, D. Lodge), a scholarship from the Illinois-Indiana Sea Grant (to JMD), an Environmental Protection Agency Science to Achieve Results Graduate Research Fellowship (to JMD), and a University of Notre Dame Schmitt Graduate Research Commission grant (to principal investigator, D. Lodge), funded by the National Science Foundation (DEB 02-13698) and the University of Notre Dame.

Competing interests. The authors have declared that no competing interests exist.

Author contributions. JMD conceived and designed the experiments, performed the experiments, analyzed the data, contributed reagents/materials/analysis tools, and wrote the paper.

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