Haze smoke impacts survival and development of butterflies

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The Southeast Asian transboundary haze contains a mixture of gases and particles from forest fires and negatively impacts people’s health and local economies. However, the effect of the haze on organisms other than humans has not yet been sufficiently studied. Insects are important members of food webs and environmental disturbances that affect insects may impact whole ecosystems. Here we studied how haze directly and indirectly affects the survival, growth, and development of insects by rearing Bicyclus anynana butterflies under artificially generated smoke as well as reared in clean air but fed on plants previously exposed to smoke. Direct haze exposure significantly increased the mortality of caterpillars, increased larval development time, and decreased pupal weight, while indirect haze exposure, via ingestion of haze-exposed food plants, also affected development time and pupal weight. No smoke particles were found in the tracheae of subjects from the smoke treatment suggesting that the increase in development time and mortality of B. anynana under smoke conditions might be due to toxic smoke gases and toxic food, rather than particulate matter. These results document significant deleterious effect of haze smoke to the development, adult size, and survival of insects, key players in food-webs.

The Southeast Asian (SEA) transboundary smoke plumes, commonly known as “haze”, occur when commercial tree plantation developers and local farmers, mainly from the Indonesian islands of Sumatra and Borneo, clear forest lands using fire1,2. The phenomenon peaks during the intermonsoon dry season between June to October and is exacerbated when there are prolonged droughts during El Nino years3-5. When carried by the monsoon winds, the haze affects local dwellers as well as those in neighboring countries such as Malaysia and Singapore6-7.

The haze impacts both human health and local economies8,9. Field studies have identified over 100 compounds in the smoke from forest fires. They include carbon dioxide and monoxide (CO2 and CO), methane (CH4), nitrogen oxides (NOX), sulphur dioxide (SO2), polycyclic aromatic hydrocarbons (PAHs), and smoke particles (also known as particulate matter, PM)10. Particulate matter less than 2.5 μm in aerodynamic diameter (PM2.5), especially, can invade the smallest airways in the lungs, aggravating respiratory and cardiovascular diseases11-13. For instance, the 1997 SEA haze episode reached hazardous levels of ground-level particulate matter concentrations, and was followed by an increase in respiratory problems in Singapore, together with a 2% annual increase in regional adult cardiovascular mortality14. In addition, haze has brought economical losses because of falls in tourism, aviation, industrial investment and production8. Despite the high number of studies on the effect of the haze on human health and lifespan, its impact on other species and on ecosystems has rarely been addressed.

Herbivorous insects are primary consumers that link plants to higher-level consumers in food webs. They are also key players in many ecological functions14, such as in plant pollination15, seed dispersal16,17, or soil aeration18. Because some insects have high sensitivity to environmental change, they are used as bioindicators of the health of a habitat or an ecosystem (e.g.19,20). Butterflies, in particular, are used for these purposes because, in addition to their sensitivity to environmental disturbances, they are easy to identify and monitor. In Indonesian forests, for instance, butterfly species richness and abundance was directly linked to human activities such as forest logging and fires, presumably via direct destruction of their habitat21,22. In addition to direct impacts, forest fires can have indirect effects on butterflies as well as other insects via the generation of smoke plumes. No study, however, has investigated the effects of smoke haze on insects.

One way in which haze might affect insects is by impacting their respiratory system, as it does for humans. Insects breath air via spiracles, valve-like openings on the side of their bodies that connect to internal tracheal tubes23,24 which branch repeatedly into finer tracheoles that eventually reach every cell inside the insect’s body,

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where diffusion of gases occurs. Insects such as flies, beetles, crickets or moths, among others, can also use an active respiration process where rapid cycles of compression and expansion of the trachea and the air sacs in various parts of the body create air movement. An insect’s aeration system constitutes, thus, a direct path to all internal organs making it extremely vulnerable to air quality. For instance, insecticide injected into spiracles leads to immediate and severe paralysis compared to a topical cuticular application.

This study aims to evaluate the effects of haze smoke on insect larval growth and survival to understand how insects might respond to SEA haze episodes. We used the nymphalid butterfly Bicyclus anynana as a test species. We hypothesized that haze has a detrimental effect on fitness, either via mechanical obstruction and damage of the fine tracheoles by the particular matters, via gas toxicity, or via food plant toxicity, when food plants alone are exposed to haze. In two sets of experiments, we exposed individuals to smoke throughout their entire development, from egg to adulthood, and in another set we exposed food plants to haze and fed individuals on these exposed plants for their entire larval development, from 1st instar to pre-pupae. We measured development time, weight at pupation, and mortality. We also performed a histological examination of the larvae trachea, and analyzed the structure of the larval spiracles via scanning electron micrographs, to gain extra insights into the mode of action of smoke on insect mortality.

Methods

Subjects and experimental setup. Eggs of Bicyclus anynana were collected from a laboratory stock reared in a climate room at 27 °C and 60% relative humidity. On the fifth day after hatching, larvae were sorted into treatment groups. The experiments were carried out outdoors, under an open-air shed, protected from direct sunlight.

The experiments were conducted inside wire mesh cages which sides were sealed by transparent PVC sheets. The bottom half of one of the sides was covered with a High Efficiency Particulate Air (HEPA) filter. To facilitate air flow inside the cage, a fan was placed outside facing the filter (Fig. 1a, b). We simulated haze using incense smoke, because of their similar composition and the incense’s ability to sustain combustion over time (one coil lasted around 24 hours). Incense is generally made of a paste of rounded wood matter along with organic adhesive and potassium nitrate to facilitate combustion. Pollutants emitted from burning incense include CO, NO, SO2, carbonyls, volatile organic compounds, PAHs, PM10 (particulate matter less than 10 µm in aerodynamic diameter), and PM2.5. In one of the cages from each treatment type, a DC1100 Pro air quality monitor (Dylos Corporation) was placed to log the daily average of smoke particle concentration (Fig. 1a, b). Temperature and humidity in the 4 cages were recorded once every day around 5 pm with a digital thermohygrometer to ensure that these conditions were constant across both treatments (Fig. 1c–e).

Direct effect of smoke on B. anynana development. We conducted two experiments (experiments 1 and 2) that differed in how the larvae were fed. In experiment 1, two cages were employed for each treatment group, Smoke and Control, making a total of four cages. Larvae were placed in groups of ten into cylindrical plastic containers, and five containers were used in each cage. Larvae were given ~0.075 g per individual of pieces of fresh young corn leaves every day. We recorded the number of surviving larvae in each container on a daily basis. In addition, we recorded the time taken by each larva to reach pupation, the number of pupae that successfully emerged, and the total mortality from egg to emergence.

In experiment 2, one cage per treatment was used. Caterpillars were reared on potted corn plants changed every one to two days, with ad libitum access to food. Larvae were placed in groups of ten per plant inside a separate mesh container, and five plants/containers were used in each cage. In addition to the same parameters described above, we also recorded the time taken by each pupae to emerge as a butterfly; the total development time from egg to adult; and the mass of each pupa on the day of pupation. Individuals that perished during the experiment were collected and kept at −20 °C for subsequent analyses.

Indirect effect of smoke-induced plant toxicity on B. anynana development. In experiment 3, we fed newly hatched caterpillars with 2-week-old potted corn plants that were either previously placed in the haze environment or in the control environment for 24 hours. As in experiment 2, the larvae were placed in groups of ten per plant inside mesh containers, with five plants used from each environment. All larvae from both treatments were kept in a control cage environment. Similar to experiment 2, we measured the time taken by each larva to reach pupation and emergence, the larval, pupal and total mortality, and the pupal mass on the day of pupation.

Histological examination of the tracheal system. We investigated the presence of smoke micro-particles in the insects’ tracheal system using light microscopy. Trachea sections from the thoracic segments (air intake trachea) were dissected in PBS and transferred to 1% glutaraldehyde in 0.1 M phosphate buffer, pH 7, for 5 minutes. Samples were then mounted onto glass slides, and observed using a Zeiss Axio Imager M2 microscope.

Spiracle structure examination using scanning electron microscopy (SEM). To investigate for the potential presence of micro particles at the spiracle openings, larval specimens were dried in an incubator at 37 °C until constant dry mass, then sputtered with a thin layer of gold using JEOL JFC 1100 ions sputtering machine, and imaged under a JEOL JSM-6510LV SEM.

Statistical analyses. Since we used only one or two cages per treatment (where each cage contained multiple groups of larvae), our results may suffer from a pseudoreplication problem (but see). We accounted for this by using mixed models, which is a statistical procedure commonly used to correct for pseudoreplication.

The effect of the smoke on the temperature, humidity and PM2.5 concentrations were analyzed in two steps. (1) since [PM2.5] was measured in only one cage of each treatment, whereas temperature and humidity were measured in 2 cages each, we analyzed the effect of the cage on these two factors separately by fitting a linear mixed model (LMM), where treatment (smoke or control) was a fixed factor, and cage (two different cages
per treatment) a random factor. The effect of the cage factor on humidity and temperature was tested using Likelihood Ratio Tests (LRTs) of full models tested against simplified models with the specific factor removed. (2) As temperature, humidity and [PM2.5] are dependent factors, we also applied a Multivariate Analysis of Variance.

Figure 1. Experimental set-up and environmental conditions inside the cages. (a,b) Individuals were grown in plastic containers or on plants (4), inside cages covered with transparent PVC sheets (7). Each smoke cage contained an incense coil (1), a metal container collecting the ashes from the coil (2), and an air quality monitor (3). A fan (5) was placed facing the bottom part of the cage on one side covered by an HEPA filter (6). The control cages’ set-up was identical, except for the incense coil and the metal container that were removed. (c,d) Temperature and humidity were similar in both control and smoke cages. (d) Particle concentration was significantly higher in the smoke cage than in the control cage. In panels c, d and e, the higher and lower bars of the plot are the maximum and minimum values respectively, while the rectangle illustrates the first quartile, the median, and the third quartile (bottom to top). The black dot is the average, and each open circle is a data point. Significant differences between treatments are represented by asterisks: *p < 0.05; **p < 0.01, ***p < 0.001, ns = non-significant.
Results

Smoke cages contained more smoke particles than control cages, but humidity and temperature were similar under both treatments. The multivariate analysis indicated that environmental conditions differed between the smoke and the control cages (MANOVA, Pillai's Trace = 0.50, $F_{1,116} = 38.12$, $p < 10^{-3}$). The average PM2.5 particle concentration was 117 µm$^{-3}$ in the smoke cage, and 50 µm$^{-3}$ in the control treatment, which represents a significant difference (ANOVA, $F_{1,116} = 111.35$, $p < 10^{-3}$) (Fig. 1d). However, temperature and humidity were similar in both environments (ANOVA, temperature: $F_{1,116} = 4.00$, $p = 0.95$; humidity: $F_{1,116} = 0.02$, $p = 0.88$) (Fig. 1e,f). Because humidity and temperature were the same across all four cages (LRT: $\chi^2 = 0$, $p = 1$), data from both cages of the same treatment in experiment 1 were pooled in subsequent analyses.

In experiment 1, a significantly higher number of larvae died in the smoke treatment, resulting in fewer individuals reaching the pupal stage, compared to the control group (GLMM: $\chi^2 = 5.89$, $p = 0.02$) (Table 1). At the pupal stage there was a large mortality in both treatment groups, and only 25% of the control (17 individuals) and 12% of the smoke treated individuals (6 animals) emerged as butterflies (Table 1). There was no difference in pupal mortality between the two groups (GLMM: $\chi^2 = 3.14$, $p = 0.08$) (Fig. 2b), but total mortality from egg to butterfly was higher in the smoke treatment (GLMM: $\chi^2 = 3.94$, $p = 0.05$) (Fig. 2c). However, smoke didn't significantly impact time to pupation (LMM; $F_{1,118} = 0.05$, $p = 0.83$) (Fig. 2d) (Table 1), and the identity of the rearing pot did not impact any of the measured traits (for all dependent variables, LRT: $\chi^2 = 0$, $p = 1$).

In experiment 2, there was high larval mortality in the presence of smoke. Overall, 5 butterflies emerged from the 50 larvae reared in the smoke cage, whereas 38 adults emerged from the 50 individuals submitted to the control environment (Table 1). Smoke strongly reduced the probability of larvae surviving to the pupal stage and to butterfly emergence (GLMM; Larval mortality: $\chi^2 = 18.76$, $p = 1.48$ $10^{-5}$; Total mortality: $\chi^2 = 16.18$, $p = 5.77$ $10^{-4}$) (Fig. 1g). Smoke did not significantly impact pupal mortality (GLMM; $\chi^2 = 0.04$, $p = 0.84$) (Fig. 1f) but larvae reared in the smoke cage took longer to reach the pupal stage compared to those reared in the control cage (LMM; $F_{1,48} = 11.72$, $p = 0.01$) (Fig. 2h; Table 1). The average time to emergence from pupation was similar in both treatments, as all butterflies emerged after 6 or 7 days (LMM; $F_{1,48} = 3.16$, $p = 0.08$) (Fig. 2i) (Table 1), but larvae reared in the smoke cage took a significantly longer time to emerge as butterflies, compared to those reared in the control cage (LMM; $F_{1,48} = 13.54$, $p = 6.90$ $10^{-4}$) (Fig. 2j) (Table 1). Pupae from the smoke cage were also significantly lighter than their control counterparts (LMM; $F_{1,48} = 18.83$, $p = 5.00$ $10^{-3}$) (Fig. 2k). Larval cohort didn't impact any of the traits measured (LRT: $\chi^2 = 0$, $p = 1$).

Smoked plants did not impact caterpillar survival but affected development time and pupal weight. Smoked plants didn't affect larval and pupal survival. Larval mortality on smoked and control plants reached 22% and 26% respectively, and all the pupae from the control plant emerged, while 2 from the smoked plants died before emergence. Therefore, 74% of the individuals reached emergence in both treatments (GLMM; larval survival: $\chi^2 = 1.17$, $p = 0.68$; pupal survival: $\chi^2 = 1.91$, $p = 0.17$; total survival: $\chi^2 = 0$, $p = 1$) (Table 1, Fig. 2l). Despite similar survival across plant treatments, caterpillars reared on smoked plants took a bit longer to reach the pupal and the adult stages than caterpillars that grew on control plants (LMM; development time to pupae: $F_{1,24} = 8.73$, $p = 4.00$ $10^{-2}$; development time to adult: $F_{1,24} = 9.56$, $p = 0.02$) (Fig. 2m) (Table 1). The pupae from individuals reared on smoked-plants were also smaller than the ones reared on control plants not exposed to the haze (LMM; $F_{1,24} = 18.22$, $p = 5.77$ $10^{-3}$) Table 1, Fig. 2o). Yet, adults reared on both types of plants took similar amount of days to emerge from the pupation step (LMM; $F_{1,24} = 0.50$, $p = 0.48$, Table 1). Finally, the larval cohort didn't affect any of the measured traits (LRT: $\chi^2 = 0$, $p = 1$).

Histological examination of the trachea. Microscopic examination of the trachea did not reveal any particle inside the respiratory tract of B. anynana larvae from the smoke or from the control group. Thoracic and abdominal sections of the trachea and tracheoles didn't contain any visible particles that could be blocking the airways (Fig. 3).

Scanning electron microscopy of spiracles. Examination of thoracic spiracles showed some particles trapped in the branched trichomes covering the spiracle's entrance valves (Fig. 4). A few particles corresponding to the sizes of PM10 and PM2.5 were observed on the spiracle trichomes of one animal from the smoke treatment group only, but we could not quantify them since they were rare.
Discussion

Incense burning successfully generated PM2.5 at concentrations that simulated real haze episodes. The World Health Organization air quality guidelines recommend concentrations of PM2.5 to be kept below 25 $\mu$g m$^{-3}$ over a 24 h period, in order to prevent damage to human health. The Singapore National...
Environment Agency considers air quality unhealthy when PM2.5 rises above 55 µg m\(^{-3}\) over a 24 h period\(^46\). The high amount of PM2.5 in the smoke cages (117 µg m\(^{-3}\)) suggested that the burning of coil incense was effective in producing unhealthy polluted air conditions, comparable to particular matter counts taken during the last haze events in Southeast Asia (SEA) countries. Haze episodes are recurrent in SEA, happening every 1 to 3 years since 1982\(^47,48\). In Singapore, the 2013 and 2015 episodes lasted around one and 2 months respectively, with [PM2.5] higher than 100 µg m\(^{-3}\) for 4 to 12 consecutive days, including 3 consecutive days when the concentration exceeded 230 µg m\(^{-3}\) in 2013\(^49–51\). In countries closer to the actual fires, the [PM2.5] remained dangerously high for long periods of time. For example, in 1997, the annual mean PM2.5 concentration reached 200 µg m\(^{-3}\) in Southeast Sumatra and Southern Borneo\(^12\). The September-October 2015 average [PM2.5] was above 300 µg m\(^{-3}\) in east Sumatra and south Kalimantan islands, Indonesia\(^52\). Human health and wildlife damages attributed to the haze are therefore expected to be extremely severe in these areas\(^12\).

Aside from differences in smoke particles measured across the two cage treatments, there were also likely differences in air composition across the two cages that were not quantified. Ambient temperature and humidity, however, were the same across all cages, so these environmental factors are unlikely to have contributed to differences in larval fitness observed across treatments.

The smoke and the smoke-induced plant toxicity impaired the development of larvae. Caterpillars that grew subjected to continuous smoke, but not the ones that were fed with smoked-plants, suffered higher mortality in both experiments 1 and 2. The individuals reared in smoke in experiment 2 also had longer development time and lower pupal mass when compared to those that grew in clean air conditions. Individuals that were fed with smoked plants also had increased development time and reduced weight, although the effect sizes were smaller than in experiment 1. Therefore, it seems that the smoke alone induces caterpillar mortality, and that both the smoke and the plant toxicity induced by the smoke increased larval development time and reduced pupal weight, two important components of fitness. These results are in line with a meta-analysis which concluded that arthropods from areas with air pollution were generally smaller than those from control sites\(^53\). As smoke-treated larvae, and larvae fed with smoked plants have a reduced pupal mass, they likely resulted in smaller butterflies with reduced fecundity compared to the control-reared ones\(^54,55\). Since PM2.5 were absent from the tracheal system of the dead caterpillars reared in the smoke cage, we cannot conclude that the impaired development was caused by particles blocking the insect’s airways. We suggest, instead, that individuals, upon detection of poor air quality or air particles, closed their spiracle valves for longer time durations, decreasing O\(_2\).
intake and reducing metabolism. Alternatively, toxic incense smoke components might have been responsible for the higher larval mortality, longer development time, and the lower pupal mass.

The individuals reared in the smoke environment and on smoked plants may have suffered from food stress. Food deprivation was previously shown to lengthen the larval development time and produce smaller butterflies, and may have originated from larvae having difficulty in locating the corn leaves, or from poor leaf quality. Pollutants such as nitrogen oxides, emitted in high quantities by burning incense and during forest fires, enhance the formation of ground-level ozone by reacting with floral or foliar volatile compounds. Ozone was shown to alter the plant volatiles that herbivores and pollinators use as cues to locate their food source, inducing them to search more and to forage less. In addition, ozone and sulfur components can both induce necrotic patches on the plant leaf surfaces and reduce the herbivore food quality. Silkworm larvae, for instance, have non-uniform growth and delayed cocooning when they are on host plants exposed to SO2. Junonia coenia caterpillars raised on foliage exposed to high-CO2 grew more slowly and experienced greater mortality, especially in early instars, than those raised on foliage exposed to low-CO2, probably due to the reduced foliar water and nitrogen concentrations of those plants. Thus, the B. anynana caterpillars reared with incense may have had difficulties locating the corn leaves, may have breathed toxic compounds and ingested more toxic food than their control counterparts, leading to slower development and smaller pupal weight. In support of the first interpretation we also observed that larvae from the smoke treatment group had larger pieces of corn uneaten after each day of feeding, than larvae in the control treatment.

The consumption of smoked plants impacted larval development time and pupal weight less than the joint effect of the smoke and plant toxicity. Thus, in addition to food stress, the larvae from experiments 1 and 2 might have also suffered from chronic inflammation. As air exchange in insects occurs directly via diffusion through tracheoles to the surrounding tissues, acidic SO2 and NO2 from the smoke may directly cause tissue damage in the larvae, initiating an inflammation. Under certain stress factors, immune response signals can reduce larval growth rates and delay molting. In mice, SO2, NO2, and PM2.5 induced inflammatory and endothelial dysfunction in the heart.

Finally, the caterpillars from experiment 1 and 2 might have suffered from hypoxia, a deficiency in the amount of oxygen necessary for the tissue to work properly. Although the proportion of gases was not measured in our experiments, oxygen levels may have been lower in the smoke cages due to incense combustion. The air blown through the HEPA filter might have not been sufficient to renew the oxygen inside the smoke cage. Hypoxia often results in the reduction of insect survival, growth rate, body size, and reproduction, while inducing the damage of macromolecules and inflammation. Insects can respond to hypoxia by decreasing their activities, such as

**Figure 4.** Scanning electron microscopy images of caterpillars and its thoracic spiracles. (a) A dried caterpillar reared in the smoke condition with a thoracic (air intake) spiracle highlighted by a white rectangle. (b) Enlarged view of thoracic spiracle. (c,d) Detail of trichomes covering the valves of the spiracles where a few micro-particles close to 2.5 µm are seen (arrow).
feeding or moving, increasing the risk of food stress. Insects can also compensate for a reduction of O\(_2\) in their environment by opening their spiracles and increasing abdominal pumping of air, processes that also increase the intake of toxic gases. Testing these inter-linked hypotheses to identify the precise cause of increased mortality and development time in caterpillars growing in smoke environments remains open for future studies.

**Smoke did not impact pupal mortality.** The similar pupal mortality observed in both treatments, and across the 3 experiments, suggests that smoke and the smoke-induced plant toxicity had little impact on pupal development. This is intriguing as there is high metabolic activity and oxygen demands during the late pupal stage especially, and the air exchange increases, including the intake of more toxic gas. In experiments 1 and 2, an insufficient sample size at the pupal stage caused by high mortality up to that point may have impaired the discovery of significant effects of smoke on pupal development. Alternatively, the low mortality at this stage might indicate that the effect of gases on food location and food quality are more important factors in reducing larval fitness than direct air toxicity, as pupae do not feed.

**Different types of food might have contributed to the different patterns of mortality and development time observed in experiments 1 and 2.** Smoke-treated larvae fed in cups (experiment 1) suffered relatively less mortality compared to the ones fed on whole plants in experiment 2 (compare Fig. 2a with e). This difference cannot be explained by food toxicity since larval mortality in experiment 3 was low. However, locating the food might have been hard for the larvae in experiment 2, especially after the plants were replaced, because individuals were placed at the bottom of the pots rather than on top of the cut leaves (as in exp. 1). In addition, both caterpillar groups had longer larval development times in experiment 1 (~35 days) relative to experiment 2 (23 and 28 days) and 3 (~22 and 24 days). This increased development in experiment 1 might have led to overall smaller pupae and to higher pupal mortality. While we did not measure the weight of pupae in experiment 1, we observed that the time to pupation and pupal weight were inversely correlated in animals of experiment 2 and experiment 3. Feeding larvae in containers also led to an overall higher pupal mortality of both groups relative to the one measured in experiment 2 and 3 (compare Fig. 2b with f). This high mortality happened as most larvae entered into the pupation process, but failed to complete it, forming an unfinished pupa that would die after one day. The longer development time and higher pupal mortality in experiment 1 were likely due to food deprivation, as the corn amount allocated to the animals was rationed. Late larval stages of both treatment groups ate all the leaves allocated to them on a daily basis. This was not the case in experiment 2 and 3 where the animals were fed ad libitum on a live plant. These results are similar to earlier studies showing that starvation at the larval stage extended *B. anynana* larval development time, reduced the larval growth rate, increased the pupal mortality and decreased the butterfly body size, fecundity and reproductive investment.

**The spiracles might be effective barriers against particulate matter (PM) and toxic gases.** The valves of spiracles may function both as mechanical filters for PM as well as sensing organs that regulate spiracle closing and air intake. The absence of smoke particles in larvae trachea and the presence of particles trapped in the trichomes covering the spiracles suggest that these structures might be effective barriers against PM entering the larval body, as previously suggested. This remains to be thoroughly tested. In addition, upon sensing particles, these trichomes might induce closure of the spiracle valve, preventing particles from entering the trachea. While we have no direct observations supporting this mechanism, mechanical disturbances are known to cause the closure of spiracles in cockroaches. Similar muscles and innervation exist in lepidopteran spiracles. Physiological studies also postulated that muscles that close the spiracles of Cecropia moth pupae are sensitive to smoke alone are the likely culprits in increasing larval mortality. This study takes a first step in understanding the impact of smoke on a population of butterflies. In natural environments, however, haze may affect multiple members of a complex food web, with less predictable outcomes. For example, in pollution-induced deteriorated areas, herbivore insects showed an increase in abundance because the density of their predators also decreased. Additional studies, where the exact composition and gas levels are precisely controlled, for instance, are needed to further understand how the smoke haze affects the equilibrium of ecosystems.

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Author Contributions

E.D. and Y.Q.T. collected and analyzed the data; A.M. and Y.Q.T. designed the study; E.D. created the figures; all authors wrote the manuscript, E.D. and A.M. contributed to revisions. E.D. and Y.Q.T. contributed equally to the manuscript.

Additional Information

Competing Interests: The authors declare no competing interests.

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