Adsorption of Basic and Acidic Dyes onto Agricultural Wastes

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Abstract. The potential of almond leaves powder, (ALP) for the removal of Crystal violet (CV) and Congo red (CR) dyes from aqueous solution was investigated. The adsorbent (ALP) was characterized by FTIR and SEM analysis. Batch adsorption studies were conducted and various parameters such as contact time, adsorbent dosage, initial dye concentration, pH and temperature were studied to observe their effects in the dyes adsorption process. The optimum conditions for the adsorption of CV and CR dyes onto the adsorbent (ALP) were found to be: contact time (100mins), pH (10.0), temperature (343K) for an initial CV dye concentration of 50 mg/L using adsorbent dose of 1.0 g and contact time (100mins), pH (2.0), temperature (333K) for an initial CR dye concentration of 50 mg/L using adsorbent dose 1.0 g respectively. The experimental equilibrium adsorption data fitted best and well to the Freundlich isotherm model for both CV and CR dyes adsorption. The maximum adsorption capacity of ALP was found to be 22.96 mg/g and 7.77 mg/g for the adsorption of CV and CR dyes respectively. The kinetic data conformed to the pseudosecond-order kinetic model. Thermodynamic quantities such as Gibbs free energy (ΔG0), enthalpy (ΔH0) and entropy (ΔS0) were evaluated and the negative values of ΔG0 obtained for both dyes indicate the spontaneous nature of the adsorption process while the positive values of ΔH0 and ΔS0 obtained indicated the endothermic nature of the adsorption and increased randomness during the adsorption process respectively for the adsorption of CV and CR onto ALP. Based on the results obtained such as good adsorption capacity, rapid kinetics, and its low cost, ALP appears to be a promising adsorbent material for the removal of CV and CR dye stuff from aqueous media.

Introduction

Colour is considered as the first pollutant to be identified in waste water and it is mostly caused by effluents discharged from dyeing industries such as paper, rubber, leather, cosmetics, textile, pharmaceuticals, plastics and food industries [1]. More than 10,000 dyes are commercially available and more than 7×105 tonnes per year of these dyes are produced in the world for the textile industry alone [2]. Most of the dyes used in the industry are synthetic; having complex aromatic molecular structures that make them more stable and less biodegradable [3]. The manufacture and use of synthetic dyes for dyeing fabrics has become an industry solid. It is estimated that about 20% of the total amount of dyes produced and used by these industries in a year is unloaded in to the industry waste without previous treatment [3].

However, their use has become a matter of serious concern to environmentalist. Synthetic dyes are highly toxic, causing negative effects on all life forms because they present sulphur, naphtol, vat dyes, nitrates, acetic acid, surfactants, enzymes chromium compounds and metals such as copper, arsenic, lead, cadmium, mercury, nickel, cobalt and certain auxiliary chemicals [4].

The crystal violet (CV) dye is a synthetic cationic dye and transmits violet colour in aqueous solution. It is also known as Basic Violet 3, gentian violet and methyl violet 10B, belonging to the group of triarylmethane. This dye is used extensively in the textile industries for dying cotton, wool, silk, nylon, in manufacture of printing inks and also the biological stain, a dermatological agent in veterinary medicine. The CV is toxic and may be absorbed through the skin causing irritation and is harmful by inhalation and ingestion. In extreme cases, CV can lead to kidney failure, severe eye irritation leading to permanent blindness and cancer. Therefore, removal of this dye from water and wastewater is highly essential [4].
Congo red (CR) is a benzidine-based azo dye with complex chemical structure, high solubility in aqueous solution and it is persistent once it is discharged into natural environment. Congo red mainly occurs in effluents discharged from textile, paper, printing, leather industries etc., and during dyeing operation; about 15% of it ends up in wastewaters [5,6]. It exists as a brownish-red crystal and is stable in air with a solubility of 1g/30ml in water. It is an anionic acid dye used as a laboratory aid in testing for free hydrochloric acid in gastric contents, in diagnosis of amyloidosis, as an indicator of pH. It has a strong affinity to cellulose fibers and thus is employed in textile industries. It is a derivative of benzidine and naphthoic acid and metabolizes to carcinogenic products. It is investigated as a mutagen and reproductive effector. It is a skin, eye and gastrointestinal irritant. It may affect blood factors such as clothing and induce somnolence and respiratory problems [6].

Diverse methods of treatment exploited through the years by researchers and industries for removing dyes and colour include physicochemical, chemical, and biological methods, such as flocculation, coagulation, precipitation, adsorption, membrane filtration, electrochemical techniques, ozonation, and fungal decolorization [7, 8]. However, due to the fact that effluents contain different dyes, and these dyes contain complex structures, is very difficult to treat using conventional methods [4].

The adsorption is one of the most effective processes of advanced wastewater treatment, which industries employ to reduce hazardous pollutants present in the effluents. This is a well-known and superior technique to other processes for removal of dyes from aqueous solutions worldwide due to initial cost, operating conditions and simplicity of design [9]. Currently, the most common procedure involves the use of activated carbon as adsorbent for this purpose by offering greater adsorption capacities. However, due to their relatively high cost, many lower-cost adsorbents have been investigated as adsorbent for removing contaminants from wastewater. The low cost adsorbents can be made from waste materials, thus collaborating with the environment and also getting economic advantages. A wide variety of low cost adsorbents have been prepared from different materials utilizing industrial, biomass, and municipal wastes [10].

These days, there are numerous low cost, commercially available adsorbents for dye removal. However, the adsorption efficiencies of these adsorbents are not very high. In view of these limitations, there have been constant search for alternative, low cost and easily available adsorbents with high efficiencies for removal of dyes from wastewater or aqueous solution. Almond (Prunus dulcis) leaves from the Almond tree constitute a nuisance as they litter the environment as waste materials. The usage of these waste materials for adsorption process could result in waste utilization and minimization leading to tremendous cost saving.

Therefore, the potential of almond leaves powder was explored under batch conditions. The effect of several parameters such as contact time, adsorbent dosage, initial dyes concentration, pH and temperature were investigated in the experimental study. The adsorption equilibrium isotherm models, kinetic models and thermodynamic parameters related with the process were studied and reported.

Literature Review

Spent tea leaves as low-cost adsorbent have been investigated by Bajpai and Jain [11] for the removal of basic dye (Crystal violet) from aqueous solutions at different temperature in batch experiments. The adsorption was found to be favoured at higher temperature. The order to fitness of various isotherm models, when applied on equilibrium uptake data, was found to be Langmuir = Temkin > Freundlich. The values of Langmuir constants, Q_max (mg/g) and b (l/mol) are found to be 114.94 and 80457.6; 185.18 and 52795.2; 285.71 and 41983.2 at 25º, 37º and 50ºC respectively. Thermodynamic quantities, i.e. ΔGº, ΔHº and ΔSº were also evaluated. The negative ΔGº values indicated spontaneous nature of the adsorption process.

Alshabanat et al. [12] examined the adsorption of Crystal violet (CV) onto date palm fibers (DPFs) in aqueous solution at 25°C. The experimental maximum adsorption capacity value was found to be 0.66×10^-6 mg/g. The percentage removal of CV dye by adsorption onto DPF at different
pH and temperatures showed that these factors play a role in the adsorption process. Thermodynamic analysis was performed, and the negative values of $\Delta G^0$ indicate spontaneous adsorption. The negative value of $\Delta H^0$ indicates that the interaction between CV and DPF is exothermic, and the positive value of $\Delta S^0$ indicates good affinity between DPF and CV. The kinetic data were fitted to a pseudo-second-order model.

Adsorption of Congo red (CR) from aqueous solution using dried roots of Eichhornia crassipes was studied by Wanyonyi et al. [13]. Batch experiments were carried out for sorption kinetics and isotherms. Experimental results obtained showed that adsorption process was highly dependent on contact time, adsorbent dosage, initial dye concentration and particle size. The sorption equilibrium for Congo red dye by E. crassipes (roots) was reached within 90 minutes and adsorption efficiency of up to 96% achieved. The sorption kinetics followed a pseudo-second-order kinetic model while Freundlich isotherm model was best applicable for obtaining the equilibrium parameters.

Nagda and Ghole [14] investigated the biosorption of congo red by hydrogen peroxide treated tendu (Diospyros melanoxylon) waste. Comparison was done on the extent of biosorption between untreated and treated forms of the adsorbents. It was reported that the adsorption occurred very fast initially and attained equilibrium within 60 minutes at pH = 6.2 and the equilibrium attained faster after modification. The data fitted Freundlich isotherm model for treated ones while it fitted Langmuir for untreated ones and the maximum adsorption capacity was 46.95 and 134.4mg/g for untreated and treated ones respectively. The kinetic data followed pseudo-second order rate equation.

Other low-cost adsorbents investigated and found efficient includes: the indian jujuba seeds which yielded adsorption capacity of 55.56 mg/g for Congo Red dye [15], jute stick powder also yielded adsorption capacity of 35.7 mg/g for Congo Red dye [16], etc.

Materials and Methods

Materials

Almond leaves powder, (ALP) (adsorbent), Congo red (CR) and Crystal violet (CV) dyes (adsorbates), Hydrochloric acid (HCl) and Sodium Hydroxide (NaOH), Distilled water, Conical flask, and beakers, Measuring cylinders and funnels, Filter papers (Whatman NO.1), Digital pH meter, Electrical shaker, Electronic weighing balance, Thermostatic water bath.

Preparation of Adsorbent

Almond leaves were collected from Naze, Owerri north local government of Imo state, Nigeria. The almond leaves were washed thoroughly with distilled water to remove sand, dirt and other impurities present. The leaves were sundried for several days until all moistures were removed. It was then grounded in a steel mill and sieved in a sieve shaker of particle size of 150µm. The almond leaves powder that passed through the sieve was used as adsorbent without any further pretreatment. It was labeled ALP (almond leaves powder) and finally stored in an airtight container for further use.

Preparation of CV and CR Dye Solutions

Analytical grade CV dye (molecular formular: $C_{25}H_{30}N_3Cl$, molecular weight: 407.98, $\lambda_{\text{max}} = 590\text{nm}$). Analytical grade CR dye (molecular formular: $C_{32}H_{22}N_6O_6Cl$ $S_2Na_2$, molecular weight: 696.68, $\lambda_{\text{max}} = 500\text{nm}$) both dyes were obtained from chemical laboratory in FUTO and used without purification. A stock solution of CV dye of concentration 1000mg/L was prepared by dissolving 1g of CV dye powder in 1L of distilled water. A stock solution of CR dye of concentration 1000mg/L was also prepared by dissolving 1g of CR dye powder in 1L of distilled water. Experimental working solutions of desired concentration were obtained by appropriate dilution of the stock solutions.
Adsorbent Characterization

Fourier transform infra-red (FTIR) spectrophotometer was used to identify the different functional groups available on the adsorbent sites. The FTIR of the adsorbent was taken before adsorption using FTIR spectrophotometer (Shimadzu-8400S). The adsorbent was encapsulated with KBr spectroscopy grade and by introducing the mix in a piston’s cell of a hydraulic pump with compression pressure 15KPa/cm²; the solid translucent disk was obtained. The FTIR spectrum was then recorded within the wave number range 4000 – 500 cm⁻¹. In addition, surface morphology and texture of the adsorbent was analyzed using scanning electron microscope (SEM) (Model-PHENOM ProX). Prior to scanning, some quantity of the adsorbent was placed on a double adhesive sticker placed in a sputter machine for 5 sec; this gave the adsorbent a conductive property. Sample (adsorbent) stud was fixed on a charge reduction sample holder, and then was charged in the SEM machine.

Batch Adsorption Experiments

Batch adsorption of CV and CR dye onto the adsorbent (ALP) was conducted in a 250ml airtight Erlenmeyer flask containing 100ml of known concentration of dye solution and an accurately weighed amount of the adsorbent. The mixtures in the flasks were agitated on a rotary shaker operating at a constant speed of 100 rpm. The effect of contact time (10, 20, 40, 60, 80, 100 & 120 min), adsorbent dosage (0.5, 1, 2, 3 & 4 g/L), initial CV dye concentration (25, 50, 75, 100 & 120 mg/l), pH (2, 4, 6, 8 &10) and temperature (303, 313, 323, 333 & 343 K) were evaluated. The flask containing the samples were withdrawn from the shaker at predetermined time intervals, filtered and the final concentrations of CV dye in the supernatant solutions were analyzed using the UV-visible spectrophotometer (Jenway 6305). The entire batch adsorption procedure was also carried out using CR dye. The pH of the solutions were adjusted using 1M HCl or NaOH. All experimental studies were carried out in triplicates and the mean values reported. The amount of equilibrium uptake of dye was determined using

\[
\text{Dye removal (\%) = } \frac{C_0 - C_e}{C_0} \times 100, \quad (1)
\]

\[
q_e = \frac{(C_0 - C_e)}{V} \times \frac{W}{M}, \quad (2)
\]

where \(q_e\) is the amount of dye taken up by the adsorbent at equilibrium (mg/g), \(C_0\) is the initial dye concentration (mg/L), \(C_e\) is the dye concentration at equilibrium (mg/L), \(M\) is the mass of the adsorbent (g), and \(V\) is the volume of the solution, (L).

Results and Discussion

Characterization of the ALP

The FTIR spectrum of ALP before adsorption was analyzed to determine the vibration frequency changes in their functional groups and is shown in Fig. 1. Various peaks were observed at 3453.34cm⁻¹ (due to O-H stretch, H- bonded), 2388.92cm⁻¹ (due to N–H bond), 2267.40cm⁻¹ (due to C≡C bond, or C≡N), 1647.26cm⁻¹ (due to C≡C bond), 1437.98cm⁻¹ (due to symmetrical aliphatic bonding of CH₃), 1187.23cm⁻¹ (due to C–N bond) 1034.84cm⁻¹ (due to C–O bond), 641.35cm⁻¹ (due to C–Cl bond), [17].

Also, Fig. 2, shows the SEM image of ALP. It can be observed from Fig. 2, that the external surface of ALP is very rough, irregular and contains pores.
Fig. 1: FTIR spectra of ALP

Fig. 2: SEM image of ALP at 1500x
Effect of Contact Time

The effect of contact time on the removal of CV and CR dyes by ALP is depicted in Fig. 3. It is observed that the extent of adsorption of both dyes was rapid at the initial stages and becomes slow in later stages until equilibrium was achieved. It can be inferred from the rapid sorption rate at the early stages that there were abundance of active sites on the external surface of ALP which resulted in the rapid CV and CR dye removal. The final CV and CR dyes concentration did not vary appreciably after 100 min from the start of adsorption process. It indicates that equilibrium was assumed to be achieved after 100 min for the adsorption of CV and CR dye by ALP. This is attributed to the saturation of the active sites on ALP adsorbent which do not allow further adsorption to take place [18].

Effect of Adsorbent Dosage

The effect of adsorbent (ALP) dosage on the removal of CV and CR dyes is depicted in Fig. 4. The result revealed that the removal efficiency of CV and CR dye increased with increase in adsorbent dosage. This observation can be attributed to the fact that the number of active sites on ALP surface increased with the increase in the adsorbent dosage which led to increase in the dyes removal [19].

Fig. 3: Effect of contact time on the percentage removal of CV and CR onto ALP
Effect of Initial Dye Concentration

The effect of initial dye concentration on the removal of CV and CR dyes is depicted in Fig. 5. It can be observed that the removal efficiency of CV and CR dye onto ALP decreased from 94.75% to 96.25% and 86.04 to 92.0% respectively as the initial concentration increased from 25mg/L to 125mg/L. This observation can be attributed mainly to the fact that, for a given mass of adsorbent, the amount of dye it can adsorb is fixed. Thus, the higher the concentration of the dye the smaller the percentage it can remove [10]. Furthermore, the equilibrium adsorption capacity increased with increase in initial dye concentration (figure not shown). This is attributed to the fact that increasing dye concentration increases the driving force to overcome all mass transfer resistances of the CV and CR between the aqueous and solid phase, leading to an increase in equilibrium adsorption capacity [19, 20].
Effect of pH

The initial pH of dye solution has greater significance than the final pH of the solution in adsorption reaction. The influence of pH on the adsorption of CV and CR onto ALP is illustrated in Fig. 6. It was observed that the percentage removal of CV dye increased while the percentage removal of CR dye decreased as the initial pH of the solution increased from 2 to 10. Optimum removal efficiency was obtained at pH of 10 and 2 for the adsorption of CV and CR dye respectively. The increase in removal efficiency of CV dye with increase in pH is basically due to the fact that at higher pH values, the positive charges (H⁺) at the solution interphase decreases and the adsorbent surface is more negatively charged, thus enhancing adsorption of more amount of the cations of the CV dye [21]. While the increase in removal efficiency with decrease in pH for the adsorption of CR can be attributed to the fact that at low pH values of the solution, the presence of excess hydrogen ion or positive charges at the solution interface increases, hence the ALP surface becomes more positively charged leading to the adsorption of more anions of CR dye.

Effect of Temperature

The influence of temperature on the removal of efficiency of CV and CR dyes by ALP is illustrated in Fig. 7. It can be observed that increase in temperature resulted in increase in removal efficiency of both CV and CR dyes. This increase in removal efficiency of CV and CR dyes by ALP with increase in temperature can be attributed to the fact that an increase in temperature may reduce the viscosity of the solution and increase the rate of diffusion of the adsorbate molecules across the external boundary layer and in the internal pores of the adsorbent (ALP) leading to more CV and CR removal [22]. Optimum adsorption was observed at temperature of 70°C and 60°C for adsorption of CV and CR dyes respectively.
Fig. 7: Effect of temperature on the percentage removal of CV and CR onto ALP

Adsorption Isotherms
Two adsorption isotherm models have been applied to describe the relationship between the CV and CR dye concentration in the bulk solution and that on the ALP’s surface namely, the Langmuir and Freundlich isotherm models.

The linear form of Langmuir equation is given as

$$\frac{C_e}{q_e} = \frac{1}{q_m K_l} + \frac{C_e}{q_m},$$

(3)

where $q_e$ (mg/g) is the amount of dye adsorbed at equilibrium, $q_m$ (mg/g), the amount of dye adsorbed when saturation is attained, $C_e$ is the equilibrium dye concentration (mg/L) and $K_l$ is Langmuir constant related to the binding strength of dye onto the adsorbent. The essential features of the Langmuir isotherm may be expressed in terms of equilibrium parameter $R_l$, which is a dimensionless constant referred to as separation factor [10,23]:

$$R_l = \frac{1}{1 + K_l C_o}$$

(4)

The value of $R_l$ indicates the type of the isotherm to be either unfavourable ($R_l > 1$), linear ($R_l = 1$), favourable ($0 < R_l < 1$) or irreversible ($R_l = 0$). The linearized form of Freundlich equation is expressed as

$$\log q_e = \log K_f + \frac{1}{n} \log C_e$$

(5)

where $K_f$ and $n$ are the Freundlich constants that represent adsorption capacity and intensity (strength) of adsorption respectively.
Fig. 8: Langmuir isotherm plot for the adsorption of CV and CR onto ALP

Fig. 9: Freundlich isotherm plot for the adsorption of CV and CR onto ALP

Fig. 8, and Fig. 9, show the Langmuir and freundlich adsorption model plot for the adsorption of CV and CR onto ALP respectively. The isotherm parameter and $R^2$ values are summarized in Table 1. Based on the values of correlation coefficient ($R^2$) (Table 1) the Langmuir plot was inadequate in describing the adsorption processes. The experimental data fitted well with the Freundlich isotherm model, hence, indicating the adequacy of the Freundlich isotherm to describe the adsorption of CV and CR onto ALP and such adsorptions mainly occurred on the heterogeneous surface of the almond leaves powder. Furthermore, Freundlich isotherm constant, $n$, (Table 1) was found to be less than 1 indicating that the adsorption of CV and CR by ALP is a chemical process [19, 24].
Table 1: Isotherm Parameters and values of correlation coefficient for CV and CR adsorption

|                      | Lamburg | Freundlich |                    |                    |
|----------------------|---------|------------|--------------------|--------------------|
|                      | q_{max} | K\_L       | R\^2               | K\_F               |
| CV onto ALP          | 22.957  | 0.072      | 0.072              | 0.803              |
| CR onto ALP          | 7.772   | 0.062      | 0.062              | 0.615              |

### Adsorption Kinetics

The Pseudo-first-order kinetic mode is represented by

$$\log (q_e - q_t) = \log q_e - \frac{k_1}{2.303} t,$$

where $q_e$ and $q_t$ are the values of amount of the dye adsorbed per unit mass on the adsorbent at equilibrium and at various time $t$, respectively, $k_1$ is the Pseudo-first-order adsorption rate constant (min^{-1}). The values of $k_1$ and calculated $q_e$ can be determined from the slope and intercept respectively, of the linear plot of log ($q_e - q_t$) versus $t$.

The pseudo-second-order kinetic model is expressed by

$$\frac{t}{q_t} = \frac{1}{k_2 q_e^2} + \frac{1}{q_e} t,$$

where $k_2$ is the pseudo-second-order adsorption rate constant (g/mg/min) and $q_e$ is the amount of dye adsorbed (mg/g) on the adsorbent at equilibrium.

The initial adsorption rate, $h$ (mg.g^{-1}.min^{-1}) is expressed as:

$$h = k_2 q_e^2$$

The plot of $t/q_t$ versus gives a linear relationship which allows computation of $k_2$, $h$ and calculated $q_e$.

Among these models, the criterion for their applicability is based on judgment on the respective correlation coefficient ($R^2$) and agreement between experimental and calculated value of $q_e$ [9,25].

![Fig. 10: Pseudo first order kinetic plot for the adsorption of CV and CR onto ALP](image-url)
Fig. 11: Pseudo second order kinetic plot for the adsorption of CV and CR onto ALP

Fig.10 and Fig 11 depict the pseudo-first-order kinetics and pseudo-second-order kinetics for the adsorption of CV and CR onto ALP. The pseudo-first-order rate constant (K1) and qe determined from the model are presented in Table 2. It is observed from Fig.10 that the relationship between the dye solution diffusivity, Log (qe – qt) against t is non-linear. It is also observed (from Table 2) that the calculated qe values did not agree with the experimental qe since the calculated qe values were neither equal nor reasonably close to the experimental qe values. Therefore, pseudo-first-order model was inadequate in representing the adsorption of CV and CR onto ALP.

The pseudo-second-order rate constant K2, and qe determined from the model are presented in Table 2. It can be observed from Table 2, that there is a good agreement between the calculated qe value and the experimental qe value i.e. qe (cal.) and qe (exp.) for the adsorption of CV and CR onto ALP. It is therefore evident that pseudo-second-order is the best fitted kinetic model and is therefore very adequate in describing and representing the kinetics of the adsorption of CV and CR onto ALP. Hence the adsorption process may be controlled by chemisorptions and it played a significant role in the rate determining step [26, 27]. This also indicates that the forming of interaction between the adsorbates (CV and CR) and adsorbent on the external surface of the adsorbent is the rate limiting step [18].

Table 2: Pseudo-first order and second order kinetic model parameters for CV and CR adsorption

|                  | Pseudo-First Order Model | Pseudo-Second Order Model |
|------------------|--------------------------|---------------------------|
|                  | qe (exp) {mg/g}          | k1 {g.mg⁻¹.min⁻¹}         | qe (cal) {mg/g} |
|                  |                          |                          | R²             |
| CV               | 4.842                    | 0.040                    | 0.149          | 0.837          | 16.76         | 0.713         | 4.850        | 0.9999        |
| CR               | 3.969                    | 0.047                    | 1.043          | 0.591          | 1.86          | 0.116         | 3.998        | 0.998          |

Adsorption Thermodynamics

The thermodynamic parameters such as change in Gibb’s free energy (ΔG°), change in enthalpy (ΔH°) and change in entropy (ΔS°) for the adsorption of the dyes onto ALP have been determined using the following equations:
The Gibb’s free energy, \( \Delta G \), equation is expressed as follows [28, 29],

\[
\Delta G = -RT\ln K_{ads}
\]  \( (9) \)

The equilibrium constants (\( K_{ads} \)) were calculated according to the following equation:

\[
K_{ads} = \frac{Dye \ concentration \ on \ the \ solid \ (adsorbent) \ at \ equilibrium (q_e^m)}{Dye \ concentration \ in \ solution \ at \ equilibrium (C_e)}
\]  \( (10) \)

The Van’t Hoff equation is expressed as follows:

\[
\ln K_{ads} = \frac{\Delta S^0}{R} - \frac{\Delta H^0}{RT}
\]  \( (11) \)

where \( R \) is gas constant, \( T \) is absolute temperature in Kelvin, \( \Delta S^0 \) and \( \Delta H^0 \) are entropy and enthalpy respectively. The plot of \( \ln K_{ads} \) versus \( 1/T \) gives a linear relationship, which allows the computation of \( \Delta H \) and \( \Delta S \) values from the slope and intercept respectively.

Fig. 12 depicts the linear Van’t Hoff equation plot for the adsorption of CV and CR onto ALP.

Negative \( \Delta G^0 \) values obtained at all temperatures (Tables 3 and 4) indicate the feasibility and the spontaneous nature of CV and CR dyes adsorption onto ALP. It can also be observed from Table 3 and 4, which the values of \( \Delta G \) for the adsorption of CV and CR onto ALP decreased with increase in temperature. This indicates a rapid and more spontaneous adsorption at the lower temperature of the adsorption process [29].

The positive \( \Delta H^0 \) values obtained for the adsorption of both dyes confirm the endothermic nature of the adsorption processes. While the positive values of \( \Delta S^0 \) reveal increased randomness at the solid/solution interfaces during the adsorption of CV and CR onto ALP [22, 26].

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**Table 3:** Thermodynamic data and parameter for CV adsorption onto ALP

| \( T \) (K) | \( \Delta G \) (KJ/mol.K) | \( \Delta H \) (KJ/mol.K) | \( \Delta S \) (J/mol.K) |
|-------------|-----------------|-----------------|-----------------|
| 303         | -4601.12        |                 |                 |
| 313         | -5064.76        |                 |                 |
| 323         | -5266.58        | 16.106          | 67.667          |
| 333         | -6456.53        |                 |                 |
| 343         | -7420.56        |                 |                 |

**Fig.12:** Van’t Hoff plot for the adsorption of CV and CR onto ALP
Table 4: Thermodynamic data and parameter for CR adsorption onto ALP

| T (K) | ΔG (KJ/mol.K) | ΔH (KJ/mol.K) | ΔS (J/mol.K) |
|-------|---------------|---------------|--------------|
| 303   | -788.312      | 10.481        | 37.340       |
| 313   | -854.91       |               |              |
| 323   | -1287.32      |               |              |
| 333   | -1404.81      |               |              |
| 343   | -2473.51      |               |              |

Conclusions

The potential of almond leaves (ALP) powder as an effective adsorbent for the removal of crystal violet (CV) and Congo red (CR) dye from aqueous medium has been identified. The adsorption of CV and CR onto ALP powder was found to be influenced by contact time, adsorbent dose, initial concentration, temperature and initial pH of the solution. The optimum adsorption of CV and CR dyes was found at pH 10 and 2 respectively. Maximum adsorption capacity was 22.957mg/g and 7.772mg/g at 50mg/L initial CV and CR dye concentration respectively. Temperature had strong influence on the adsorption processes and the maximum removal was observed at 343K and 333K for CV and CR dyes respectively. The equilibrium was attained at 100min for the adsorption of CV and CR dyes. The kinetic studies revealed that the adsorption of CV and CR dyes onto ALP followed the pseudo-second-order kinetic model. The study on equilibrium sorption revealed that Freundlich isotherm model gave the best fit to the experimental data. The calculated thermodynamic parameters such as negative ΔG and positive ΔH values obtained indicated a spontaneous and endothermic nature of the adsorption of CV and CR dyes onto ALP powder respectively. While the positive ΔS values obtained for both adsorption of CV and CR indicated increase in randomness during the adsorption process. The present results confirm the use of ALP powder as an inexpensive and efficient adsorbent without any pretreatment or modification for the removal of Crystal violet and Congo red dyes from aqueous solutions and waste effluents.

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