Safeguarding Imperiled Biodiversity and Evolutionary Processes in the Wallacea Center of Endemism

MATTHEW J. STRUEBIG, SABHRINA G. ANINTA, MARIA BEGER, ALESSIA BANI, HENRY BARUS, SELINA BRACE, ZOE G. DAVIES, MAARTEN DE BRAUWER, KAREN DIELE, CILUN DJAKIMAN, ROSIE DRINKWATER, ALEX DUMBRELL, DARREN EVANS, MARCO FUSI, LEONEL HERRERA-ALSINA, DJOKO T. ISKANDAR, JAMALUDDIN JOMPA, BERRY JULIANDI, LESLEY T. LANCASTER, GINO LIMMON, LINDAWATI, MICHAELA G. Y. LO, PUNGKI LUPIYANINGDYAH, MOLLY MCCANNON, ER IK MEJUAARD, SIMON L. MITCHELL, SONNY MUMBUNAN, DARREN O’CONNELL, OWEN G. OSBORNE, ALEX S. T. PAPADOPULOS, JOENI S. RAHAJOE, ROSARIA, STEPHEN J. ROSSITER, RUGAYAH, HIMMAH RUSTIAMI, ULRICH SALZMANN, SHEHERAZADE, I MADE SUDIANA, ENDANG SUKARA, JOHY S. TASIRIN, AIYEN TJOA, JUSTIN M. J. TRAVIS, LIAM TRETHOWAN, AGUS TRIANTO, TIM UTTERIDGE, MARIA VOIGT, NURUL WINARNI, ZULIANTO ZAKARIA, DAVID P. EDWARDS, LAURENT FRANTZ, AND JATNA SUPRIATNA

Islands offer fascinating insights into the world’s evolutionary processes. Perhaps nowhere is this more evident than the archipelagos of Wallacea, where the Asian and Australasian biogeographic regions collide. Wallacea, which is made up of the islands of Sulawesi, the Moluccas, and Lesser Sunda in Indonesia and Timor Leste, supports the highest levels of endemism worldwide (Mittermeier et al. 2011), including 62 endemic vertebrate genera and iconic oddities such as the babirusa (Babyrousa spp.), the maleo (Macrocephalodon spp.), and the Komodo dragon (Varanus komodoensis). Recognizing that the fauna was so strikingly different from that of neighboring Borneo and Papua, nineteenth century naturalist Alfred Russel Wallace changed the course of evolutionary theory by delineating what ultimately became a distinct biogeographic region (figure 1). Although the bounds of Wallacea have, at times, included other islands, it is the central transition zone between the Sahul and Sunda land masses in Indonesia that has long captured scientific attention (Lohman et al. 2011, Ali and Heaney 2021).

With new marine and terrestrial species discoveries still occurring across the region, the importance of Wallacea as a treasure trove of biodiversity is increasingly evident (Rheindt et al. 2020, Esselstyn et al. 2021). Nonetheless, Wallacea’s ecosystems face considerable pressure from a growing human population (33.7 million people in 2021; www.bps.go.id). Economic development and people’s livelihoods have largely centered on the exploitation of the region’s natural resources, but the consequences of mining, forestry, agriculture, hunting, tourism, and fisheries for biodiversity and particularly for the endemic species that make

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Wallacea so unique are concerning. In the present article, we review the historic and contemporary processes shaping Wallacea’s unique ecosystems and explore ways to conserve its hyperdiversity.

Wallacea: A natural laboratory for the study of evolution
The abrupt biogeographic transitions of the Wallacea fauna, as is demarcated by Wallace’s line to the west and Lydekker’s line to the east, are the result of long-term isolation of the archipelago from Asia and Australasia, respectively (Lohman et al. 2011, Ali and Heaney 2021). Faunal migration and speciation were shaped by a complex geological history, involving rifting of the Makassar Strait during the Eocene, collisions of Sunda margins with the Australian continent during the Miocene, and tectonic movement during the early Neogene (Hall 2013). Although most of the small islands in the southeast, as well as Timor and Seram, emerged around 5 million years ago, the largest island, Sulawesi, originally formed as a series of smaller land masses that only amalgamated into its current form around 1 million years ago (Vaillant et al. 2011, Hall 2013, Nugraha and Hall 2018). This merger created distinct volcanic ridges, mountains, and ancient lakes that characterize the island today. Although the formation of Sulawesi as a single island is relatively recent, deep ocean trenches to the east and west have separated it, as well as the wider Wallacea region, from the Sunda and

Figure 1. Maps of Wallacea showing its location in Southeast Asia (a), zones of endemism on land (b) and in coastal marine environments (c). See supplemental tables S1 and S2 for further information on the zones. In panel (b), the shading represents Indonesia’s primary forest cover in 2018 compiled from Margono and colleagues (2014) and data from www.globalforestwatch.org. Forest according to this definition cannot be reproduced consistently for Timor-Leste, so forest is not displayed for this country. Forests are also difficult to map consistently in the drier parts of Wallacea, notably the Lesser Sunda islands (Nusa Tenggara); see box 1. In panel (c) mangrove cover (green/dark shading) is derived from Bunting and colleagues (2018) and corals (orange/light shading) from https://allencoralatlas.org.
Sahul continental shelves throughout the past 23 million years (Hall 2013).

This complex geological history and subsequent diverse environments makes Wallacea a natural laboratory for testing the roles of dispersal, geographic isolation, and adaptation in driving evolution, speciation, and ecological diversification. The evolutionary history of the many endemic species that are found in the region’s terrestrial and freshwater ecosystems (figure 2) illustrate well how the rise and fall of geological barriers can promote the evolution of species (supplemental tables S1 and S2). The ancestors of many endemic lineages arrived long after the region became isolated from Asia and Australia in the Eocene (Stelbrink et al. 2012), indicating that migration through rafting and island hopping likely played a strong role in the assembly of Wallacea’s unique terrestrial ecosystems. There are now more than 120 endemic fishes (e.g., Telmatherininae, sail-fin silversides) and invertebrates (e.g., Tylomelania snails and Syntripsa crabs) that have subsequently diversified in Sulawesi’s ancient lakes, and colonization and isolation in montane regions have triggered further speciation in mammals (Crocodida), frogs (Limnonectes), and plants (Cyrtandra; table S2). Although Wallacea covers less than 2% of the world’s oceans, its marine ecosystems contain approximately 76% of the world’s hard coral species and more than 2000 shallow noncryptic coral reef fishes (Veron et al. 2009).

In some taxa, populations and species likely diversified in isolation on separate paleoislands, perhaps with occasional dispersal across the paleoarchipelago. Subsequent periods of large-scale uplift over the past 4 million years (Nugraha and Hall 2018) had further impacts on biodiversity by connecting previously separated regions (Frantz et al. 2018). Indeed, this uplift may have reunited previously isolated species and subspecies of terrestrial taxa, including tarsiers, shrews, macaques, snails, frogs, lizards, crickets, and damselflies (supplemental table S3), generating distinctive phylogeographic patterns (figure 3) and some generalized zones of terrestrial endemism (figures 1 and 2, table S1; Michaux 2010). However, there are many lineages that buck these trends. For example, many freshwater lineages (e.g., Tylomelania snails Caridina shrimps, Gecarcinucid crabs and Telmatherinid fish) have undergone adaptive radiations in the island’s ancient lake systems (von Rintelen et al. 2012), and plant diversification (e.g., Cyrtandra and Nepenthes) may be more heavily influenced by soil types, elevation, and other ecological barriers (Bouchenak-Khelladi et al. 2015).

Wallacea’s complex geological history, combined with its location and the extent of its coastline at the confluence of the Pacific and Indian Oceans, also promoted the evolution
of exceptionally biodiverse coastal marine ecosystems (Tittensor et al. 2010, Wicaksono et al. 2017). The archipelagos of the region lie in the center of the Coral Triangle, the global epicenter for marine diversity, including corals, fishes, and mollusks (Veron et al. 2009). Different biogeographical hypotheses have been proposed for how this high biodiversity came to be (Veron et al. 2009, Huang et al. 2018). The region’s unstable tectonic environment continually creates diverse shallow habitats. During past sea level changes, these often remained connected to deep water, creating refuges that formed both a buffer from extinction and opportunities for divergent evolution of endemic species. The consistent movement of water and eddies created by the Indonesian throughflow ocean current increases the chance of subsequent larval dispersion at different rates across geological timescales (e.g., Linsley et al. 2010).

Wallacea’s paleontological record bears witness to these complex evolutionary processes with large numbers of island endemics (figure 2). Although less is known about Wallacea’s fossil record compared with that of neighboring Java, the oldest terrestrial fauna recorded (approximately 2.5 million years ago) from South Sulawesi includes giant suids (Celebochoerus), giant tortoises (Geochelone atlas), and two dwarf elephants (Elephas celebensis and Stegodon sompoensis; van den Bergh et al. 2001). Most of these megafauna became extinct at the end of the Pleistocene or the early Holocene. The paleontological record is also rich in hominid (other than Homo sapiens) fossils and artifacts, which first appeared in the Pleistocene around 1 million years ago (van den Bergh et al. 2001). Given that the region was never connected via a land bridge to neighboring continental shelves, these hominids would have had to complete sea crossings across treacherous waters to reach Wallacean islands with limited technologies, most likely from Borneo (Shipton et al. 2020). Wallacea was likely inhabited by at least two early hominid species, including Homo floresiensis and, potentially, Homo erectus (van den Bergh et al. 2001), both of which disappeared by the end of the Pleistocene, together with other megafauna (Sutikna et al. 2018). Archeological evidence for the presence of modern humans (Homo sapiens) indicated that they may have lived on Wallacean islands around 50,000–30,000 years ago (Shipton et al. 2020), with evidence of their activities including some of the earliest prehistoric cave paintings dated to at least 40,000 years ago (Brumm et al. 2021).

The activities of hominins, including modern humans, likely had a strong impact on the region for thousands of years. Even prior to the introduction of farming, at least 40 taxa were thought to have been translocated across Southeast Asia to Wallacean islands, including game species such as deer, pig, buffalo, and junglefowl (Heinsohn 2003). These translocations likely increased with the arrival of modern humans in the archipelago (Kealy et al. 2017), probably for hunting stock or ceremonial purposes. The process further intensified with the introduction of domesticated taxa, including pigs, dogs, and chickens, which then became feral during the expansion of farming cultures from mainland Asia (Piper 2017).

Contemporary human pressures

More than 160 years after Wallace wrote of the rich marine diversity of Ambon, the coast of this Moluccan island has been dredged, overfished, and polluted (Limmon and Marasabessy 2019). This provides just one example of the multiple stressors that have affected Wallacean environments in recent decades. But the remoteness of Wallacea, especially its eastern islands, has spared much of the region from the severe overexploitation of marine, freshwater, and terrestrial ecosystems characteristic of other parts of Southeast Asia. For instance, the rugged interior forests of most Wallacean islands remain relatively intact (Voigt et al. 2021), having avoided the pervasive conversion to large-scale plantations typical of western Indonesia, Malaysia, and southern Thailand. Indeed, between 2001 and 2019 the annual deforestation rate across Wallacea (0.39%) was half of
that of neighboring Borneo (0.76%) and a quarter of that of Sumatra (1.52%; figure 1; Gaveau et al. 2022), reflecting the continued dominance of small-scale farming and agroforestry over large-scale extractive industries (figures 4 and 5). Compared with industrial agriculture, these low-intensity farmlands support high species diversity, especially when close to forest and where habitat heterogeneity and canopy cover is maximized (Waltert et al. 2011). Nonetheless, forest specialists and endemic species are disappearing from these communities, often being replaced by wide-ranging generalists following a process of biotic homogenization across the archipelago (Maas et al. 2009, Mitchell et al. 2022).

A deforestation surge over the last 10 years, primarily in Central Sulawesi and northern Maluku, has been linked to the expansion of mining and industrial oil palm plantations (Supriatna et al. 2020). Although the region’s oil palm industry is at an early stage, mineral extraction has rapidly expanded since 2010 following a new mining governance regime in Indonesia. This led to more than 95% of the country’s nickel coming from Wallacean islands (mainly Sulawesi and Halmahera), and around half of its gold (figure 4). Elsewhere, key agricultural commodities, such as coconut and cacao (figure 4), have resulted in small-scale but widespread encroachment of forest areas (Supriatna et al. 2020), whereas mangroves have been replaced by commercial shrimp farms (Richards and Friess 2016, Malik et al. 2017).

Land-cover changes also leave Wallacea’s islands particularly susceptible to ecological invasion. Human-mediated establishment of invasive species can occur by either facilitating dispersal across islands after natural arrival or introducing those species from elsewhere. For example, genetic analyses suggest seven amphibian and reptile taxa that are now widespread in human settlements and croplands of the Lesser Sundas colonized these islands relatively recently (less than 100 years ago; Reilly et al. 2019). Invasion of the poisonous Asian common toad (Duttaphrynus melanostictus) is a key concern because the species has benefited from the absence of toads in the region and expanded quickly throughout. The large snakes and varanid lizards at the top of Wallacea’s food chains are particularly susceptible...
to poisoning from eating this prey, because, as naive predators, they are incapable of processing the toad’s toxins (Reilly et al. 2017). Sulawesi’s lake ecosystems have also been adversely affected by introduced species, including flowerhorn cichlids, Nile tilapia, and common carp (Serdiati et al. 2021). For example, in Matano, one of South Sulawesi’s ancient lakes, adult flowerhorn cichlids prey on native fish and invertebrates endemic to this ancient lake (Hilgers et al. 2018). In Central Sulawesi’s Lake Poso, this predation and competition is so severe that the number of nonnative fish species (17) now exceeds the native fish fauna (13; Herder et al. 2022).

The infrastructure development that often accompanies deforestation has facilitated hunting and wildlife trade throughout Wallacea, which is challenging to counter in more remote islands, where patrols, communication, and enforcement are limited. Wild meat is routinely sold in markets and supermarkets in North Sulawesi, for example, and around 500 metric tons of bats are imported to the province annually from across the island (Sheherazade and Tsang 2015). Egg collection of Sulawesi’s famed maleo birds (Macrocephalon maleo) has led to a rapid abandonment of nesting grounds and the species elevated to Critically Endangered status by the IUCN (www.iucnredlist.org/species/22678576/194673255). In Maluku, trade in endemic birds for the pet and songbird trade is prolific, having caused an 80% crash in salmon-crested cockatoo (Cacatua moluccensis) populations in a decade (www.iucnredlist.org/species/22684784/93046425). Because Wallacea is characterized by relatively few large-bodied terrestrial vertebrates, unsustainable extraction levels disproportionately affect a small set of these highly threatened species (Scheffers et al. 2019).

Landscape change and overexploitation have important ramifications for Wallacea’s aquatic ecosystems. Hydroelectric dams and mining pose huge threats to the integrity and functioning of freshwaters, culminating in Sulawesi ancient lakes (von Rintelen et al. 2012). A gold rush in North Sulawesi, for example, led to elevated mercury concentrations in nearby rivers and high bioaccumulation in fish (Limbong et al. 2003). Runoff from agricultural practices and urbanization has reduced downstream water quality in Ambon Bay to the extent that the sediment load smothers corals and has shifted reef communities toward sediment-tolerant species (Limmon and Marasabessy 2019). Pollution of the marine environment tends to be stronger close to major population centers, such as Makassar and Manado in Sulawesi, which also attract substantial fishing pressure. In contrast, the eastern islands of Maluku are less exposed to commercial fisheries, but artisanal fishing is widespread, and its impacts are poorly understood. Many of these islands still host the highest reef fish biomass in Indonesia (Campbell et al. 2020). However, broadscale, destructive fishing practices are evident even in remote reefs, leaving a long-term legacy of damaged reef structure and overexploited populations (e.g., overfishing of endangered Banggai cardinalfish, Pterapogon kauderni, for the aquarium trade, www.iucnredlist.org/species/63572/12692964).

Climate change further threatens Wallacean biodiversity, with at least two recent coral bleaching and mortality events degrading coral habitats, fisheries, and, therefore, income opportunities (Moore et al. 2017). The reefs around Buton and Spermonde islands off Sulawesi, for instance, show a community shift from a dominance of branching corals (e.g., Acropora spp.) toward more heat-tolerant species (Yusuf and Jompa 2012). Climate change also pushes terrestrial species outside of their thermal optima, changing phenology, and seasonality that could drive transitions toward savanna (Siyum 2020), making dry forest ecosystems of the Lesser Sunda islands (e.g., Sumbawa, Flores; box 1) particularly vulnerable. In the absence of adaptation or behavioral and phenological changes, species will have to move to track their thermal envelopes, and much of that movement on Wallacea’s rugged and convoluted islands is expected to be up slope (Harris et al. 2014). Although this high landscape complexity provides potential refuge areas that may help buffer some species against the most adverse impacts of climate change (Trew and Maclean 2021), the high numbers of species endemic to single islands or small island chains mean that the potential for mountaintop extinctions is high should these refuges prove insufficient. This is particularly concerning for species with limited dispersal capabilities or those highly dependent on intact habitat or water resources on drier and more seasonal islands in the east. These issues point toward the need for integrated land and coastal management to retain connectivity of protected forests, mangroves, and coral reefs across thermal gradients.

Wallacea at a crossroads
Sustainable use of natural resources is needed to ensure biodiversity, evolutionary processes, and ecosystem functioning are maintained while accounting for rapidly growing human pressures and accelerating climate change. The situation is further exacerbated in Wallacea by incentivizing development policies that encourage unsustainable practices in agriculture, mining, and fisheries, and therefore, the continued overexploitation of terrestrial and marine ecosystems. On one hand, low-intensity cultivation tends to have fewer major impacts on the local environment relative to large-scale agriculture (Waltert et al. 2011), and deforestation has been much slower in Wallacea than in other regions as a result (Voigt et al. 2021). On the other hand, the income potential of these livelihoods is often insufficient, making some parts of Wallacea (e.g., Maluku, Gorontalo in Sulawesi, and East Nusa Tenggara) among the poorest in Indonesia, whereas those that are better oriented to market-based livelihoods (e.g., North and South Sulawesi) prosper (www.bps.go.id; figure 4). Forested landscapes are often poorly suited to farming, with limited transportation and infrastructure networks, leaving residents facing high costs to access markets, education, and healthcare (Angelsen et al. 2014). Coastal communities face similar challenges with accessing
Box 1. Characterizing the forests of Wallacea.

Mapping and monitoring forests is essential for countries to meet their sustainability commitments, such as conserving biodiversity and ecosystem processes and tackling climate change (e.g., Nadin et al. 2019). But monitoring the fate of Wallacea's forests is surprisingly difficult, even with advances in remote sensing. In 2018, the islands of Sulawesi, Maluku, and Nusa Tenggara represented 141,080 square kilometers (km$^2$) of forest according to Global Forest Change data, a reduction of 10,231 km$^2$ since 2000 (Voigt et al. 2021). However, like any mapping assessment (e.g., Austin et al. 2019), this coverage remains prone to error because of ongoing challenges in distinguishing forest from regrowth and forest-like vegetation, particularly agroforestry (e.g., Sulawesi, Halmahera, and Seram; figure 5a, 5b). This is compounded in the islands of southern Maluku and Nusa Tenggara where the tropical moist forests characteristic of equatorial Southeast Asia become much drier seasonal ecosystems (figure 5c). Because seasonal forests have different spectral properties to tropical forests, their extent is likely underestimated in large-scale mapping assessments that rely solely on standard forest definitions (Staver et al. 2011).

The islands of the Lesser Sundas (i.e., Nusa Tenggara) lie within a climatic zone that could support both dry forest and savanna (Staver et al. 2011). The flora are tolerant to drought and the fauna are somewhat adapted to the phenology of these predominantly deciduous forests. Unlike savanna, seasonal forests of Nusa Tenggara include a closed canopy and no grass layer (figure 6d; Pennington et al. 2018). Baseline taxonomic knowledge of Wallacea's seasonal forests lags far behind similar ecosystems in the American tropics, where plant diversity is known to rival that of Amazonia's prominent wet tropical forests (Pennington et al. 2018). Current estimates suggest around 300 plant species endemic to the Lesser Sundas can be found in Nusa Tenggara’s seasonal forest, but this list continues to grow (e.g., Sunarti et al. 2022).

Key threats to Wallacea’s seasonal forests include mining and encroachment from small-scale agriculture (Austin et al. 2019, Voigt et al. 2021), as well as highly localized collection of firewood, timber, plants and wildlife (figure 5e, 5f). Although in tropical Southeast Asia agricultural plantings tend to be long-term cash crops, the low rainfall of the Lesser Sunda islands limits irrigation. Seasonal forests therefore tend to be replaced by rotating cultivation systems, which, over time, may return to forest vegetation (Monk et al. 1997). Unfortunately, the dry local climate makes seasonal forests highly vulnerable to these activities. A single stressor—such as fire used to clear land—can rapidly alter the stable state of forest to that of savanna. Although both savanna and seasonal forest plants can cope with drought, differences in fire regimes between these two ecosystems mean they support fundamentally different species (Pellegrini et al. 2021). Seasonal forest plants tend to be highly intolerant to fire, meaning that any switches to savanna are less reversible, posing a great risk in the highly threatened seasonal forests of Nusa Tenggara.

Figure 5. Forest and forest uses in Wallacea. (a, b) Agroforestry practiced in community managed forest of Seram, Maluku raises the question of how to define a forest from aerial imagery. (c, d) Seasonal forests in Sumbawa, Nusa Tenggara are mostly deciduous and fragmented by small-scale agriculture. (e, f) Examples of the many uses of forests for local communities include timber extraction (in the present figure, Seram, Maluku) and subsistence hunting (Flores, Nusa Tenggara). Photographs: Aris Santaya (a), Ulat Ifansasti (b, e) and Aulia Erlangga (c, f), courtesy of CIFOR.org, and Gemma Bramley (d), Royal Botanic Gardens Kew.
markets and social facilities, particularly on isolated islands. These considerations, combined with challenging law enforcement, mean that solutions to Wallacea’s emerging biodiversity crisis are not as simple as lobbying for further habitat protection.

Currently, only 9% of Wallacea’s land and 3% of its marine zone are formally protected for biodiversity conservation (figure 1; www.protectedplanet.net/country/IDN), although Indonesia has ambitious targets for protected area expansion by 2030. Similarly, although formally protected terrestrial reserves have curbed deforestation in endemic-rich areas (Voigt et al. 2021), they are limited in Buru and Flores, and many of the smaller islands. Protected areas, particularly marine ones, face many logistic, managerial, and funding challenges to be effective (Kamil et al. 2017). Moreover, expanding strictly protected areas places substantial burdens on natural resource users and are often viewed as threatening livelihoods, despite, for instance, the need for no-take zones to support sustainability of fisheries (Mills et al. 2010). It is therefore imperative to work with local communities when establishing protected areas or other conservation interventions. Much of the current conservation activity in Wallacea is focused on protected areas and where international nongovernmental organizations operate. Diversifying conservation resources, as well as increasing them, will help empower grassroots organizations and multiple stakeholders so that conservation could be more locally led.

Community-based management has a long history on land and at sea and, in Indonesia, is manifested as highly localized customary adat practices. Since 2015, these customary rights have been embodied within Indonesia’s social forestry program and extended to other local communities, regardless of cultural or ethnic background. Among the multiple social forestry schemes available, those that permit some limited timber extraction and agroforestry for local use can help improve social welfare while minimizing deforestation (Rakatama and Pandit 2020). However, uptake across the Wallacean region has been slow relative to the western islands of Indonesia (Santika et al. 2019). Correspondingly, marine community-based management (e.g., Territorial Use Rights in Fisheries) typically allows some resource extraction and can achieve sustainable fish biomass. Culturally embedded conservation practices of the sea or land in Maluku pertain to local dynamic closures that prohibit and allow fishing. Contextualizing these practices within traditional marine tenure and belief systems—sasi—has been effective in reducing overfishing (Halim et al. 2020).

Community involvement and local leadership are also vital for restoring ecosystem function and diversity. Some of the best examples come from the coastal–marine zone, where communities have been instrumental in restoring reefs and mangroves, albeit after many failed attempts (box 2). On land, Indonesia also pioneered the use of ecosystem restoration licenses through which degraded forests are leased by the government for restoration-compatible business development (Harrison et al. 2020). Licenses are granted typically for 60–100 years, allowing activities such as conservation, ecotourism, management of nontimber forest products, and tree protection for carbon sequestration. Once legally harvestable timber volumes have been restored and environmental safeguards met, a license holder may log the forest. Around one-third of the forest eligible for ecosystem restoration licenses (4,325,649 hectares by 2020. http://phil.menlhk.go.id/tabular) are in the Wallacea region, with most of these located in Sulawesi and Maluku.

Conclusions and future research directions

There remains immense potential for emerging scientific and technological tools to further document, monitor, and help prioritize Wallacea’s unique and imperiled biodiversity for conservation actions. For example, combining traditional morphological analyses with bioacoustic, genetic, and genomic data led to the description of 12 new bird taxa from Sulawesi’s satellite islands, more than double the number typically described per year globally (O’Connell et al. 2019, Rheindt et al. 2020). Molecular methods, including environmental DNA, and large-scale collaborations to catalog the genetic code of thousands of organisms are also facilitating a revolution in biodiversity monitoring. This is particularly notable in coastal ecosystems where microbial surveys in sediment and water help assess their roles in ecosystem functioning (DiBattista et al. 2020) and conservation (Bani et al. 2020). The science of restoration is also being served by population genomic and meta-omic approaches, which can be used to reveal new genotypes and the potential for genetic biocontrol of pest populations as well as for monitoring purposes (Breed et al. 2019). It is now possible to construct highly resolved food webs, moving beyond the study of specific plants or animals, toward a holistic understanding of direct and indirect species interactions that links biodiversity with ecosystem functioning (O’Connell et al. 2022).

As well as taking full advantage of scientific advances, there needs to be a fundamental shift away from managing marine, freshwater, and terrestrial realms independently. This can lead to inefficiencies and inaction, particularly where interactions and threats occur between ecosystems, such as the influence of land-use change on coastal sediment exposure (Beger et al. 2010). Taking a ridge-to-reef approach should lead to lower opportunity costs and improved outcomes for conservation in Wallacea. To be future-proof, land-use and conservation planning should also ensure these ecosystems and the endemic biodiversity they support are climate resilient (Struembig et al. 2015, Dixon et al. 2021).

Scenario planning with multiple, diverse groups of stakeholders, where a range of what-if future land- and marine-use scenarios are explored, can be helpful in identifying and understanding different viewpoints, key knowledge gaps, and the nature of interactions between inputs and outcomes (e.g., do tipping points exist beyond which ecological recovery is unlikely?). Although research can only approximate complex and nonlinear interactions, it can help to avoid inappropriate policy or practice decisions. Other sources of
Wallacea has among the highest mangrove, seagrass and coral biodiversity and biomass globally (Veron et al. 2009, Alongi et al. 2016), providing extensive ecosystem services. But these “blue” ecosystems face unprecedented pressures as aquaculture and farmland replace mangroves, disruptive fishing techniques and tourism affect seagrass meadows and reefs, and ocean warming kills corals. These pressures have affected commercially important fisheries, increased erosion, decreased coastal stabilization, and reduced water quality (Camp et al. 2016), substantially degrading valuable ecosystem services.

Mangrove restoration programs involving monospecific plantations often fail because of a poor choice of foundation species or lack of hydrological restoration. However, in Wallacea, the government, businesses, nongovernmental organizations, and local communities have had some success in restoring blue ecosystems. For example, in Bunaken National Park, local communities broke down the walls of abandoned shrimp ponds and dug trenches to reestablish tidal flow, supporting natural recruitment of mangrove propagules and associated organisms (figure 6; Djamaluddin et al. 2019). The resulting mixed-species regeneration, facilitated by the hydrological interventions (i.e., “ecological restoration”) supported diverse faunal communities and interaction networks that more closely resembled old-growth areas than monoculture plantings of similar age (O’Connell et al. 2022).

There are similar coral reef restoration success stories for the region. Novel community-based reef restoration efforts in Spermonde, South Sulawesi (c). Old growth (d) and degraded (e) mangrove forests in North Sulawesi, followed by community-based hydrological restoration (f)—that is, a channel that was reopened to connect the old shrimp pond area to the sea. Photographs: Maarten De Brauwer (a, b), Dominic Muenzel (c), Karen Diele (d, e), Rignolda Djamaluddin (f).

Future restoration needs to be implemented at a much larger scale across Wallacea to support the targets of the UN 2021–2030 Decade of Ecosystem Restoration. However, this requires broad consensus across different sectors of society, to help remove barriers such as tenure issues with other land and sea users. Upscaling coral reef and seagrass restoration to the levels required to counteract Indonesia’s bomb-fishing legacy and climate change mortality is infeasible, making prevention of ecosystem degradation a better overall strategy than restoration once the damage is done. Indeed, protected natural capital assets will always be better, both ecologically and economically, than restoration. Complementary protocols for long-term monitoring and evaluation of restoration outcomes, including biodiversity, ecosystem functioning, and livelihood benefits, are also needed.
knowledge, such as views on traditional management practices, can also help steer landscape or seascape planning in a direction where people with the greatest role in implementing sustainable practices (i.e., Wallacea’s vast and growing rural communities) are given a voice regarding managing their local ecosystems. Indonesia’s national planning processes now incorporate scenario tools, and there are moves to expand this to subnational development planning, starting with Sulawesi (Nadin et al. 2019).

Because poverty remains a major driver of unsustainable natural resource use in Wallacea, it should be tackled head on. We need a much better grasp of what evidence and which decisions underpin optimal outcomes for community wellbeing, ensuring that socioeconomic gains from the region’s development are not undermined by socioecological losses. For example, the oil palm industry, still in its infancy in Wallacea, has largely helped improve living standards elsewhere in Indonesia but has also resulted in social conflict and pollution in places (Santika et al. 2021). There is huge scope for Indonesia’s mining boom to follow the same trajectory if environmental and social safeguards are not adequately followed. Although the government requires all mining permit holders to follow good mining principles and plan for postmining reclamation before they can begin operations, land reclamation is ecologically challenging and mostly oriented toward restoring soils and basic vegetation (Pratiwi et al. 2021). Minimizing the ecological footprint and engaging nearby communities in the first place (e.g., by optimizing land-use planning using the High Conservation Value approach, www.hcvnetwork.org) is clearly a more cost-effective strategy to minimize the impacts of development and maximize the benefits it brings (Budharta et al. 2018). Improved monitoring and evaluation of mining and postmining operations are needed, to help identify best management practices that can be promoted broadly across the country. After all, Wallacea’s central role producing the nickel needed to help the global transition to low carbon technologies should bring investments that benefit local communities, and not the land disputes and environmental damage reported from some sites (Hidayana et al. 2020).

As part of these efforts, community-based and locally led approaches that have been effective elsewhere in Indonesia (Santika et al. 2019) should be further promoted in Wallacea, with communities empowered to derive sustainable livelihoods from these schemes and contribute to conservation. For instance, terrestrial and marine wildlife tourism is already an important source of income for Wallacean communities on Komodo island (Ardiantiono et al. 2018) and in North Sulawesi (Towoliu 2014). In Bangka Island, North Sulawesi, the combined actions of local communities and tourism operators successfully stopped illegal mining activities in favor of sustainable jobs and subsistence fishing (Kalalo 2019). However, tourism is not without challenges as excessive diving also affects reef conditions (Towoliu 2014). Likewise, a key need for Wallacea’s terrestrial and marine community-managed areas will be to evaluate sustainable extraction rates and integrate these appropriately in management plans alongside incentives and rewards for sustainable activities. For example, maleo numbers experienced a population recovery in Tompotika, Central Sulawesi, after villages were engaged by local conservationists and received community benefits (financial and nonfinancial) in exchange for ceasing egg poaching (Tasirin et al. 2021).

Although ecosystem restoration licenses appear to be a promising tool, the high costs of restoration have yet to be sufficiently offset via harvestable products, carbon markets, or other payments for ecosystem service schemes (Harrison et al. 2020). The success of ecosystem restoration licenses could be improved if ecosystem services (e.g., carbon, biodiversity, water) were adequately valued and the payments were allocated fairly. Ensuring that payments reach people closest to and most dependent on ecosystem services remains politically challenging. However, Indonesia’s government has taken steps in the right direction through its recognition of traditional land and natural resource use rights, and a willingness to transfer management rights from the state to rural communities (Rakatama and Pandit 2020). Clear rights and appropriate rewards may provide people with the security that prevents short-term overexploitation of resources to protect longer-term benefits. Without such changes in thinking, the environmental costs of natural resource management in Wallacea, as it does elsewhere in the world, will ultimately negate most socioeconomic gains.

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Supplemental material
Supplemental data are available at BIOSCI online.

References cited
Ali JR, Heaney LR. 2021. Wallace’s line, Wallacea, and associated divides and areas: History of a tortuous tangle of ideas and labels. Biological Reviews 96: 922–942.
Alongi DM, et al. 2016. Indonesia's blue carbon: A globally significant and vulnerable sink for seagrass and mangrove carbon. Wetlands Ecology and Management 24: 3–13.

Angelén A, Jagger P, Babigumira R, Belcher B, Hogarth NJ, Bauch S, Börner J, Smith-Hall C, Wunder S. 2014. Environmental income and rural livelihoods: A global comparative analysis. World Development 64: S12–S28.

Ardianto, Jessop TS, Purwandana D, Ciofi C, Jeri Imansyah M, Panggur AR, Sondita MFA. 2020. Transforming traditional management into contemporary territorial-based fisheries management rights for small-scale fisheries in Indonesia. Marine Policy 116: 103923.

Bail H. 2013. The paleoecogeography of SundaLand and Wallacea since the Late Jurassic. Journal of Limnology 72: e1.

Barkan, Menyuring, Kembang, Jono, Veloso, Sondita MFA. 2021. Mangrove Watch: A new 2010 global baseline of mangrove extent. Global Environmental Change 52: 152–161.

Bourdineau, Frey, Carvalho, von Rintelen. 2011. Biogeography of the Indo- Australian archipelago. Annual Review of Ecology, Evolution, and Systematics 42: 205–426.

Brum A, et al. 2021. Oldest cave art found in Sulawesi. Science Advances 7: eabf4648.

Buddhista S, et al. 2018. Restoration to offset the impacts of development at a landscape scale reveals opportunities, challenges and tough choices. Global Environmental Change 52: 152–161.

Bunting P, Rosenqvist A, Lucas RM, Rebelo L-M, Hiliarides L, Thomas N, Hardy A, Ioth T, Shimada M, Finlayson CM. 2018. The Global Mangrove Watch: A new 2010 global baseline of mangrove extent. Remote Sensing 10: 1669.

Camp EF, Suggett DJ, Gendron G, Jompa J, Manfrino C, Smith DJ. 2016. Mangrove and seagrass beds provide different biogeochemical services for corals threatened by climate change. Frontiers in Marine Science 3: 52.

Campbell SJ, Darling ES, Paredes S, Ahmadia G, Mangubhai S, Amkililiela, Estradivari, Maire E. 2020. Fishing restrictions and remoteness deliver conservation outcomes for Indonesia's coral reef fisheries. Conservation Letters 13: e12698.

DeBattista JD, Reimer JD, Stat M, Reimer JD, Masucci GD, Biondi P, De Brauwer M, Camp EF, Suggett DJ, Gendron G, Jompa J, Manfrino C, Smith DJ. 2016. Mangrove and seagrass beds provide different biogeochemical services for corals threatened by climate change. Frontiers in Marine Science 3: 52.

Campbell SJ, Darling ES, Paredes S, Ahmadia G, Mangubhai S, Amkililiela, Estradivari, Maire E. 2020. Fishing restrictions and remoteness deliver conservation outcomes for Indonesia's coral reef fisheries. Conservation Letters 13: e12698.

DeBattista JD, Reimer JD, Stat M, Reimer JD, Masucci GD, Biondi P, De Brauwer M, Wilkinson SP, Chariton AA, Bunce M. 2020. Environmental DNA can act as a biodiversity barometer of anthropogenic pressures in coastal ecosystems. Scientific Reports 10: 8365.

Dixon AM, Forster PM, Beger M. 2021. Coral conservation requires ecological climate-change vulnerability assessments. Frontiers in Ecology and the Environment 19: 243–250.

Djamaluddin R, Brown B, Iii R. 2019. The practice of hydrological restoration for coral threatened by climate change. Frontiers in Marine Science 6: 8365.

Estradivari, Maire E. 2020. Fishing restrictions and remoteness deliver conservation outcomes for Indonesia's coral reef fisheries. Conservation Letters 13: e12698.

Fayle TM, eds. Advances in Ecological Research, vol. 62: Tropical artiodactyls driven by recent geological events. Proceedings of the Royal Museum of Natural History 454: 1–108.

Fujita R, Adhuri DS, Hordyk Frantz LAF, et al. 2018. Synchronous diversification of Sulawesi’s iconic Djamaluddin R, Brown B, Iii R. 2019. The practice of hydrological restoration for coral threatened by climate change. Frontiers in Marine Science 6: 8365.

Halc T, Carvalho GR, von Rintelen T. 2011. Biogeography of the Indo- Australian archipelago. Annual Review of Ecology, Evolution, and Systematics 42: 205–426.

Harrison R, Swinfield T, Ayat A, Dewi S, Sufali M, Heriansyah I. 2020. Restoration concessions: A second lease on life for beleaguered tropical forests? Frontiers in Ecology and the Environment 18: 567–575.

Heinsohn R. 2003. Animal translocation: Long-term human influences on the vertebrate zoogeography of Australasia (natural dispersal versus ethnophoresy). Australian Zoologist 32: 351–376.

Herder F, et al. 2022. More non-native fish species than natives, and an invasion of Malawi cichlids, in ancient Lake Poso, Sulawesi, Indonesia. Aquatic Invasions 17: 72–91.

Hilgers I, Herder F, Hadiaty RK, Pfänder J. 2018. Alien attack: Trophic interactions of flowerhorn cichlids with endemics of ancient Lake Matano (Sulawesi, Indonesia). Evolutionary Ecology Research 19: 575–590.

Huang D, Goldberg EE, Chou LM, Roy K. 2018. The origin and evolution of coral species richness in a marine biodiversity hotspot. Evolution 72: 288–302.

Hudayana B, Suharko, Widyantara AB. 2020. Communal violence as a strategy for negotiation: Community responses to nickel mining industry in Central Sulawesi, Indonesia. Extractive Industries and Society 7: 1547–1556.

Kalalo FP. 2019. Revisiting the mining controversy of Bangka Island, Minahasa Regency, North Sulawesi, Indonesia. Paper presented at the 1st Warrmadewa Research and Development Seminar (WARDS), 30 October 2018, Denpasar-Bali, Indonesia.

Kamill KA, Hallu A, Rogers A, Pandit R. 2017. An assessment of marine protected areas as a marine management strategy in Southeast Asia: A literature review. Ocean and Coastal Management 145: 72–81.

Kealy S, Louys J, O’Connor S. 2017. Reconstructing palaeogeography and inter-island visibility in the Wallacean archipelago during the likely period of Sahul colonization, 65–45,000 years ago. Archaeological Prospective 24: 259–272.

Khodabakhshi, Ghahreman, Ali, Aise, Majid, Morteza, Darya, Ali, Anis. 2020. Soft coral: A new entry in the diversity of corals of northern Sistan and Baluchestan Province, south-eastern Iran. Pakistan Journal of Marine Sciences 19: 335–341.

Limpong D, Kumampung J, Rimper J, Arai T, Miyazaki N. 2003. Emissions and environmental implications of mercury from artisanal gold mining in north Sulawesi, Indonesia. Science of the Total Environment 302: 227–236.

Limmon GV, Marasabessy AM. 2019. Impacts of sedimentation on coral reefs in Inner Ambon Bay, Indonesia. IOP Conference Series: Earth and Environmental Science 339: 012035.

Lindsey BK, Rosenthal Y, Oppo DW. 2010. Holocene evolution of the Indonesian throughflow and the western Pacific warm pool. Nature Geoscience 3: 578–583.

Lohman DJ, de Bruyn M, Page T, van Rintelen K, Hall R, Ng PKL, Shih H-T, Carvalho GR, van Rintelen T. 2011. Biogeography of the Indo- Australian archipelago. Annual Review of Ecology, Evolution, and Systematics 42: 205–426.

Maas B, Putra DD, Waltiert M, Clough Y, Tscharnkte T, Schulze CH. 2009. Six years of habitat modification in a tropical rainforest margin of Indonesia do not affect bird diversity but endemic forest species. Biological Conservation 142: 2653–2671.

Malik A, Mertz O, Fensholt R. 2017. Mangrove forest decline: Consequences for livelihoods and environment in South Sulawesi. Regional Environmental Change 17: 157–169.

Margono BA, Potapov PV, Turubanova S, Stolle F. 2014. Primary forest cover loss in Indonesia over 2000–2012. Nature Climate Change 4: 730–735.

Michaux B. 2010. Biogeology of Wallacea: Geotectonic models, areas of endemicism, and natural biogeographical units. Biological Journal of the Linnean Society 101: 193–212.
Scheffers BR, Oliveira BF, Lamb I, Edwards DP. 2019. Global wildlife trade.

Santika T, Wilson KA, Law EA, St John FAV, Carlson KM, Gibbs H, Richards DR, Friess DA. 2016. Rates and drivers of mangrove deforestation.

Reilly SB, Stubbs AL, Karin BR, Arida E, Iskandar DT, McGuire JA. 2019. The origins and arrival of the earliest domestic animals.

Rakatama A, Pandit R. 2020. Reviewing social forestry schemes in Indonesia. Opportunities and challenges. Forest Policy and Economics 111: 102052.

Pennington RT, Lehmann CER, Rowland LM. 2018. Late Cenozoic paleoecography of Sulawesi, Indonesia. Palaeogeography, Palaeoclimatology, Palaeoecology 490: 191–209.

O’Connell DP, Kelly DJ, Lawless O, N’Briend K, Marcaigh FO, Karya A, Analuddin K, Marples NM. 2019. A sympatric pair of undescribed white-eye species (Aves: Zosteropidae: Zosterops) with different origins. Zoological Journal of the Linnean Society 186: 701–724.

O’Connell DP, et al. 2022. Assessing mangrove restoration practices using species-interaction networks. Restoration Ecology 30: e13546.

Pellegrini APA, et al. 2021. Decadal changes in fire frequency shift community and functional traits. Nature Ecology and Evolution 5: 504–512.

Pennington RT, Lehmann CER, Rowland LM. 2018. Tropical savannas and dry forests. Current Biology 28: R541–R545.

Piper PJ. 2017. The origins and arrival of the earliest domestic animals in mainland and island Southeast Asia: A developing story of complexity. Pages 251–273 in Piper PJ, Matsumura H, Bulbeck D, eds. New Perspectives in Southeast Asian and Pacific Prehistory, vol. 45. Springer.

Pratiwi, et al. 2021. Managing and reforesting degraded post-mining landscape in Indonesia: A review. Land 10: 658.

Rakatama A, Pandit R. 2020. Reviewing social forestry schemes in Indonesia: Opportunities and challenges. Forest Policy and Economics 111: 102052.

Reilly SB, Wogan GOU, Stubbs AL, Arida E, Iskandar DT, McGuire JA. 2017. Toxic toad invasion of Wallacea: A biodiversity hotspot characterized by extraordinary endemism. Global Change Biology 23: 5029–5031.

Reilly SB, Stubbs AL, Karin BR, Arida E, Iskandar DT, McGuire JA. 2019. Recent colonization and expansion through the Lesser Sundas by seven amphibian and reptile species. Zoologica Scripta 48: 614–626.

Rheindt FE, Prawiradilaga DM, Ashari H, Gwee CY, Lee GWX, Wu MY, Ng NSR. 2020. A lost world in Wallacea: Description of a montane archipelagic avifauna. Science 367: 167–170.

Richards DR, Friess DA. 2016. Rates and drivers of mangrove deforestation in Southeast Asia, 2000–2012. Proceedings of the National Academy of Sciences 113: 344–349.

Sanitka T, et al. 2019. Heterogeneous impacts of community forestry on forest conservation and poverty alleviation: Evidence from Indonesia. People and Nature 1: 204–219.

Sanitka T, Wilson KA, Law EA, St John FAV, Carlson KM, Gibbs H, Morgans CL, Ancrenza M, Meijaard E, Struurbek MJ. 2021. Impact of palm oil sustainability certification on village well-being and poverty in Indonesia. Nature Sustainability 4: 109–119.

Scheffers BR, Oliveira BF, Lamb I, Edwards DP. 2019. Global wildlife trade across the tree of life. Science 366: 71–76.

Serdiati N, Arfati D, Sri Wisodo M, Leono TD, Ndobe S, Mamsyur K, Moore AM. 2021. Perspectives on sustainable management of the Poso Lake (Indonesia) endemic ricefish, Oryziasanginemas (Actinopterygii: Adrianichthyidae). Revista de Biología Tropical 69: 139–152.

Sheherazade, Tsang SM. 2015. Quantifying the bat bushmeat trade in North Sulawesi, Indonesia, with suggestions for conservation action. Global Ecology and Conservation 3: 324–330.

Shipton C, O’Connor S, Kealy S. 2020. The biogeographic threshold of Wallacea in human evolution. Quaternary International 574: 1–12.

Siyum ZG. 2020. Tropical dry forest dynamics in the context of climate change: Synthesizes of drivers, gaps, and management perspectives. Ecological Processes 9: 25.

Staver AC, Archibald S, Levin SA. 2011. The global extent and determinants of savanna and forest as alternative biome states. Science 334: 230–232.

Stelbrink B, Albrecht C, Hall R, von Rintelen T. 2012. The biogeography of Sulawesi revisited: Is there evidence for vicariant origin of taxa on Wallace’s “anomalous island”? Evolution 66: 2252–2271.

Streubel MJ, Fischer M, Gaveau DLA, Meijaard E, Wich SA, Conner N, Sykes R, Wilting A, Kramer-Schadt S. 2015. Anticipated climate and land-cover changes reveal refuge areas for Borneo’s orang-utans. Global Change Biology 21: 2891–2904.

Sunarti S, Rugayah, Low YW, Lucas EJ. 2022. Syzygium nusatenggaraense (Myrtaceae), a new rainforest tree species with a calyptrate calyx from the Lesser Sunda Islands, Indonesia. Telopoea 25: 1–6.

Supriatna J, Shekelle M, Fuad HAH, Winarni NL, Dwiyahreni AA, Farid M, Mariati S, Margules C, Prakoso B, Zakaria Z. 2020. Deforestation on the Lesser Sunda Islands of Sulawesi and the loss of primate habitat. Global Ecology and Conservation 24: e01205.

Sutikna T, Tocheri M, Faith T, Jatmiko, Awe R, Meijer H, Wahyu E, Roberts R. 2018. The spatio-temporal distribution of archaeological and faunal finds at Liang Bua (Flores, Indonesia) in light of the revised chronology for Homo floresiensis. Journal of Human Evolution 124: 52–74.

Tasirin JS, et al. 2021. Malo Macrocephalon maleo population recovery at two Sulawesi nesting grounds after community engagement to prevent egg poaching. Global Ecology and Conservation 28: e01699.

Thorhaug A, et al. 2020. Longevity and sustainability of tropical and sub-tropical restored seagrass beds among Atlantic, Pacific, and Indian Oceans. Marine Pollution Bulletin 160: 111544.

Tittensor DP, Mora C, Jeltz W, Lotze HK, Ricard D, Berghe EV, Worm B. 2010. Global patterns and predictors of marine biodiversity across taxa. Nature 466: 1098–1101.

Towolli R. 2014. Coral reef condition in several dive spots around Bunaken Island, North Sulawesi. Aquatic Science and Management 2: 44–48.

Trew BT, Maclean IMD. 2021. Vulnerability of global biodiversity hotspots to climate change. Global Ecology and Biogeography. 30: 768–783.

Vaillant JJ, Haffner GD, Cristescu ME. 2011. The ancient lakes of Indonesia: Towards integrated research on speciation. Integrative and Comparative Biology 51: 634–643.

van den Bergh G, de Vos J, Sondaar PY. 2001. The late quaternary palaeogeography of mammal evolution in the Indonesian Archipelago. Palaeogeography, Palaeoclimatology, Palaeoecology 171: 385–408.

Veron JEN, Devantier LM, Turak E, Green AL, Kininnmonth S, Stafford-Smith M, Peterson N. 2009. Delineating the coral triangle. Galaxea, Journal of Coral Reef Studies 11: 91–100.

Voigt M et al. 2021. Emerging threats from deforestation and forest fragmentation in the Wallacea centre of endemism. Environmental Research Letters. 16: 094048.

Von Rintelen T, Von Rintelen K, Glaubrecht M, Schubart C, Herder F. 2012. Aquatic biodiversity hotspots in Wallacea: The species flocks in the ancient lakes of Sulawesi, Indonesia. Pages 290–315 in Gower D, Johnson K, Richardson J, Rosen B, Ruber L, Williams, S, eds. Biotic Evolution and Environmental Change in Southeast Asia, Systematics Association Special Volume Series. Cambridge University Press. doi:10.1017/CBO9780511735882.014

Waltert M, Robo KS, Kaupa S, Montoya ML, Nsanyi MS, Fermon H. 2011. Assessing conservation values: Biodiversity and endemicity in tropical land use systems. PLOS ONE 6: e16238.
Matthew J. Struebig and Sabhrina G. Aninta contributed equally as joint first author of the manuscript. Matthew J. Struebig (m.j.struebig@kent.ac.uk) is affiliated with the Durrell Institute of Conservation and Ecology, at the University of Kent, in Canterbury, England, in the United Kingdom. Sabhrina G. Aninta (s.g.aninta@qmul.ac.uk) is affiliated with the School of Biological and Behavioural Sciences at Queen Mary University of London, in London, England, in the United Kingdom. Maria Beger (m.beger@leeds.ac.uk) is affiliated with the School of Biological and Marine Science at the University of Leeds, in Leeds, England, in the United Kingdom. Maija K. Gustafsson (maia.ogren@kth.se) is affiliated with the Stockholm Environment Institute, in Stockholm, Sweden, in the United Kingdom. Daniel Pauly (daniel.pauly@ubc.ca) is affiliated with the Research Center for Ecosystems and Evolutionary Biology at the National Research and Innovation Agency, in Cibinong, Indonesia. Jatna Supriatna (jsupriatna@sci.ui.ac.id) is affiliated with the Research Center for Climate Change, at the University of Indonesia, in Depok, Indonesia. Endang Sukara (endangsukara@gmail.com) is affiliated with the Research Center for Climate Change, University of Indonesia, in Depok, Indonesia. Marinus Johannes W. Joppa (mj.joppa@gmail.com) is affiliated with the School of Biological and Behavioural Sciences at Queen Mary University of London, in London, England, in the United Kingdom. Sabhrina G. Aninta (s.g.aninta@qmul.ac.uk) is affiliated with the School of Biological and Behavioural Sciences at Queen Mary University of London, in London, England, in the United Kingdom. Jatna Supriatna (jsupriatna@sci.ui.ac.id) is affiliated with the Research Center for Climate Change, University of Indonesia, in Depok, Indonesia.