Contributions and Challenges of High Throughput qPCR for Determining Antimicrobial Resistance in the Environment: A Critical Review

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Abstract: Expansion in whole genome sequencing and subsequent increase in antibiotic resistance targets have paved the way of high throughput qPCR (HT-qPCR) for analyzing hundreds of antimicrobial resistance genes (ARGs) in a single run. A meta-analysis of 51 selected studies is performed to evaluate ARGs abundance trends over the last 7 years. WaferGen™ SmartChip is found to be the most widely used HT-qPCR platform among others for evaluating ARGs. Up till now around 1000 environmental samples (excluding biological replicates) from different parts of the world have been analyzed on HT-qPCR. Calculated detection frequency and normalized ARGs abundance (ARGs/16S rRNA gene) reported in gut microbiome studies have shown a trend of low ARGs as compared to other environmental matrices. Disparities in the HT-qPCR data analysis which are causing difficulties to researchers in precise interpretation of results have been highlighted and a possible way forward for resolving them is also suggested. The potential of other amplification technologies and point of care or field deployable devices for analyzing ARGs have also been discussed in the review. Our review has focused on updated information regarding the role, current status and future perspectives of HT-qPCR in the field of antimicrobial resistance.

Keywords: AMR; high throughput qPCR; ARGs; MGEs; gut microbiome

1. Introduction

Antimicrobial resistance (AMR) is a major public health safety issue and is not a new phenomenon in health and agricultural settings [1,2]. Recently AMR has gained more attention due to ever increasing discharge of pollutants (including antibiotics) into environmental matrices [3–5]. Detection of bacterial resistance by conventional methods is time consuming and laborious. Molecular techniques can detect antimicrobial resistance genes (ARGs) in a rapid and sensitive manner. ARGs encode the ability in bacteria to resist and grow in the presence of antibiotics. Although, ARGs presence inside bacteria pose a serious threat but their existence in environmental matrices is of equal concern because
environmental bacteria and/or clinical pathogens can acquire ARGs via horizontal gene transfer [6,7]. ARGs have been regarded as environmental pollutants in the past [8]. Previously scientists were only targeting a small fraction of ARGs in environmental matrices [9,10] but with the reduction in the price of sequencing and consequent expansion in whole genomic sequencing of bacteria [11], the number of available ARG targets in different databases have reached to thousands (Table S1). A comprehensive profiling of ARGs by conventional qPCR is time consuming, cost ineffective and burdensome. High throughput qPCR (HT-qPCR) is a relatively rapid and convenient method for the simultaneous evaluation of a large number of ARGs.

Comprehensive analysis of different ARGs is essential to decipher the intricate relationship between ARGs, MGEs and other complex and overlapping environmental factors. For example Sun and co-workers have studied effect of temperature on the ARGs during anaerobic digestion of dairy manure [12]. Similarly assessment of public health safety risks associated with exposure to antimicrobials, ARGs and/or antibiotic resistant bacteria is another important factor to fully understand this area. Insights into the role of different environmental matrices in the dissemination and/or development of antibiotic resistance is also pushing the scientific community towards high throughput profiling of resistance genes [13,14].

Application of manure and its impact on soils is of particular interest in agricultural settings, because soil microbiomes have been known to contain a diverse pool of ARGs, including the most common types of resistances found in other environmental matrices [15–17]. Presence of ARGs, without any antimicrobial exposure or treatment, in various animals and human beings is not surprising and has already been observed. For example transmission of microbial communities and associated ARGs between healthy mother-infant pairs [18], where resistance genes were detected in the mouth of new-born immediately after birth [19]. Similarly the presence of ARGs in remote pristine environments has also been reported [20]. Moreover numerous studies have evaluated the ARGs via HT-qPCR in aquatic environments indicating easy transport of ARGs and obvious association of pollution sources in rivers and estuaries [21].

Although various research articles have been published about the detection and prevalence of ARGs in various environmental matrices all over the world, there has not been any review article highlighting the role of HT-qPCR for analyzing the antimicrobial resistance genes. Therefore, the aim of our review is to provide a snapshot of the contributions made by HT-qPCR technology with the perspective of antimicrobial resistance. The specific objectives of the study are as follows: (i) to evaluate and compare HT-qPCR technology with other competitive molecular methods, (ii) to identify the global trends of antimicrobial resistance by meta-analysis of the included studies and (iii) to examine and highlight major discrepancies and disparities in the data analysis of HT-qPCR. Standardized reporting of ARGs abundance in the literature is also suggested. In the last section the potential and a path forward for future applications of new and emerging methods for evaluating ARGs have been discussed.

2. Methodology and Meta-Analysis

For analyzing the role of HT-qPCR in the field of antimicrobial resistance, we have used the Web of Knowledge database (http://apps.webofknowledge.com.proxy2.cl.msu.edu) to collect publications. Our search terms included (high throughput qPCR AND antibiotic resistance). Only studies published up to October, 2018 were included for the analysis. Research articles were then checked manually to remove any duplicates and/or irrelevant articles. Articles which are not published in English language were not considered for the present study. Eventually 51 research articles reporting the use of HT-qPCR for analyzing ARGs met our criteria and were selected for the meta-analysis. Approximately 1000 samples of different environmental matrices, analyzed in 51 included studies, were from different regions of China, United States of America, Finland, Netherlands, UK, Australia, Canada, Antarctica and Ecuador (Figure 1). In particular, 14 published reports evaluated ARGs and MGEs in soil samples from different agricultural sites and farms. Many studies have evaluated effects of chemicals, manure
and composts on the abundance of ARGs in soils. Aquatic matrices seems to receive more attention (20/51) than soil related studies.

Figure 1. World map showing no. of samples of different environmental matrices analyzed by HT-qPCR over last 7 years in different regions of the world.

The number of samples, primers, unique genes detected and normalized gene abundances (ARGs/16S) were provided either in the main text or have been retrieved from figures and/or supplementary material (Table S2). For some studies the ARGs were normalized per bacterial cells. In such cases the values have been multiplied to 4.1 in order to convert the abundances per 16S rRNA gene assuming 4.1 16S rRNA genes were present in one bacteria [22,23]. Region based distribution of samples belonging to different environmental matrices were evaluated and depicted by a world map. World map used in this review was made with the help of geo-coordinate module of heatmapper software [24]. Basic data calculations, bar graphs and pie-chart was made in Microsoft Excel 2010. Box and whisker plots for comparing the normalized ARGs abundances in different matrices were made using Paleontological Statistics Software Package for Education and Data Analysis (PAST: v3.20) [25]. Additionally the number of reference sequences in different resistance associated databases were also retrieved from the published manuscripts and/or websites of the databases (Table S1).

3. Competitor Technologies of HT-qPCR for High Throughput Profiling of ARGs

Tremendous amount of work has been published regarding the use of conventional hybridization array technology in the field of antimicrobial resistance [26–29]. The major advantage of microarray technology over HT-qPCR is the number of ARGs (~1000’s) which can be profiled in one run. However, microarray suffer batch to batch variability and generally considered as relatively less sensitive and specific. This is further exemplified by the fact that microarray data needs additional validation by qPCR [30,31] however, HT-qPCR data does not need additional validations.

Similarly large number of studies are using metagenomic sequencing approach for evaluating the status of ARGs in bacterial communities [32–35]. The major advantage of metagenomic profiling is its
ability to analyze unknown sequences whereas with HT-qPCR it is necessary to know the sequences beforehand for designing the primers. HT-qPCR, on the other hand, certainly provides better detection limits than by metagenomic sequencing approach. For example HT-qPCR has reportedly detected ARGs to the order of magnitude of $10^{-4}$ ARGs/16S rRNA gene [36,37]. Average bacterial genome size is ~3.67 to ~5.56 million bases [38] and the size of 16S rRNA gene is ~1500 bp [39]. Assuming 10,000 reads are needed to detect a single copy of 16S rRNA gene, the same detection limit for any ARG as shown by HT-qPCR would require at least $10^8$ reads during metagenomic analysis. In addition to high detection limits, the learning curve during the data analysis of HT-qPCR is less steep as it does not involve complex bioinformatics tools and pipelines, a necessity in case of short gun metagenomic approach.

Both metagenomic sequencing and microarray hybridization generally requires more sample and/or DNA quantity. In some instances, scientists have solved this problem by whole genome amplification but this can have impacts on the quality of the final results [40,41]. HT-qPCR has countered this limitation because it can perform reactions in nanoliter scale thus consuming only a minute amount of DNA. The development of HT-qPCR platforms has indeed changed the perception about conventional qPCR and its limitless utilities. This techniques has proved to be very cost-effective as nanoliter scale reactions allow substantial savings in terms of consumables and reagents and also allows a more efficient use of the available sample. All these features along with previous strongholds of conventional qPCR (sensitivity and specificity) have made HT-qPCR a technique of better choice in the context of AMR detection.

HT-qPCR technology has some disadvantages as well. For example, inability of HT-qPCR to optimize individual assays during a run as all assays would experience same qPCR cycling conditions. This can be critical because specific binding of primers often require different annealing temperatures [42,43]. Also the reactions at nanoliter scale make it difficult to recover and sequence amplified products which is easily possible in conventional PCR.

**Different HT-qPCR Platforms Used for ARGs Analysis**

Currently four different kind of HT-qPCR platforms have been used for profiling of ARGs. These platforms have different high throughput capacities and also uses different reaction volumes (Figure 2). All these platforms have been used for measuring the diversity and abundance of ARGs in different environmental and clinical samples but WaferGen’s platform is the most frequently used platform for analysis and profiling of ARGs. In 75% of the studies included in our analysis, WaferGen™ Smartchip has been used for analysis of antibiotic resistance (Figure 2).

![Figure 2. Number of publications using different HT-qPCR platforms over last 7 years.](image)
Size of the reaction volume can affect qPCR results [44]. Bio-Rad CFX384™ platform provides better analytical sensitivity due to its larger reaction volumes but because of a relatively smaller number of reaction wells (384), it has been employed only in two ARGs related studies for high throughput profiling of the ARGs [45,46]. On the other hand very high throughput capacity (96×) of Biomark™ Dynamic Array (Fluidigm, California, CA, USA) have forced the reduction of reaction volumes to 7–9 nL which can negatively affect its analytical sensitivity. Also the required pre-amplification step [47,48] can also effect the accuracy of the analysis. Open Array® of Applied Biosystems™ allows a higher reaction volume (33 nL) with a reasonable throughput capacity (32×) but the primers in this case has been pre-dispensed by the company causing inflexibility in terms of run to run customization (Figure 3).

WaferGen™ Smartchip provides customized no. of samples versus assays formats which can be opted according to the requirements of the experimental design. Most of the studies have used 296 versus 18 and 384 versus 12 Assay-Sample format. It also has better analytical sensitivity as compared to other microfluidic chips because of larger 100 nL reaction volume. It is perhaps due to these features that scientists are preferring the use of WaferGen™ Smartchip over others HT-qPCR platforms especially in the field of antimicrobial resistance.

4. Role and Impact of Community Dynamics and MGEs in ARGs Dissemination

Changes in resistome of bacterial population in response to shifts in environmental parameters such as temperature, pH and so forth, can occur either through vertical gene transfer (selective enrichment of a bacterial population) [49,50] or via horizontal gene transfer of ARGs [51]. Increase in abundance of ARGs along with changes in bacterial community composition and subsequent enrichment of any particular bacterial group have been widely reported. For example Enterobacteriaceae increase was positively correlated with ARGs in lakes in Brazilian Zoo [52]. Also during high throughput profiling of ARGs in East Tiaoxi river, role of bacterial community was found to be more effective as compared to mobile genetic elements in influencing the resistome [53]. Similarly, Xiang and colleagues, while studying the spatial and temporal distribution of antibiotic resistomes indicated the variations in bacterial community structure was major contributing factor in the alterations of bacterial resistomes [54]. However, this does not diminish the importance of horizontal gene transfer, an important pathway for dissemination of ARGs, in the development and emergence of multidrug

![Figure 3. Reaction Volume and High throughput capacity of various HT-qPCR platforms used for ARGs analysis.](image-url)
resistant bacteria in different clinical and environmental matrices [55]. For example, sub-lethal stress caused by environmental fluctuations such as pH downshifts is reported to accelerate the net rate of inter- and intra-specific transmission of selected antimicrobial resistance plasmids [56]. Jacquiod and colleagues have assessed the conjugative plasmid permissiveness of wastewater microbiomes and found that Gammaproteobacteria were preferred hosts highlighting the role of plasmid in mobility of ARGs via horizontal gene transfer [57].

In silico and experimental analysis of MGEs was performed in several studies in order to understand their role in transportation and/or enrichment of ARGs in different environmental matrices [58,59]. Int1 gene is known as a major player in the transport of ARGs both within and between different bacterial species. Several molecular assays involving int1 gene has already been developed to get an overall view of the ARGs as they are believed to be a proxy of anthropogenic pollution and ARGs [60,61]. Many studies have implied that the co-occurrence of MGEs especially int1 can play a negative role by disseminating ARGs in clinically important strains, thus can produce multiple antibiotic resistance bacterial strains which can cause serious public health and safety concerns [62–64].

Except few studies included in our analysis, most studies analyzing ARGs have also evaluated the status of MGEs. A total of 9 MGEs including int1 were present in 296 primer array. This array has been used in 68.6% (35/51) of the studies included in the review. The number of MGEs was increased to 61 in 384 primer array that has been utilized in various studies [36,65]. In more than half of the studies (26/51), researchers have also analyzed the community structure by sequencing the hypervariable regions of 16S rRNA gene. Different hypervariable regions were amplified and sequenced in different studies, which can affect and/or influence the results of community analysis [66,67]. We will not be discussing the impact and selection criteria for amplification and sequencing of hypervariable regions of 16S rRNA gene as it is out of the scope of the review.

In (27/51) studies researchers have reported significant linear correlation between certain ARGs and MGEs including int1. However, correlation of ARGs with MGEs or int1 is not evident enough to infer that horizontal gene transfer is responsible for enrichment of ARGs. As integrons allow acquisition and dissemination of ARGs within gene cassettes [59,68]. Possibility of an altogether increase of ARGs and MGEs, if the cell carrying the particular plasmid has been positively selected, is always there. In that case microbial community dynamics would be the major driver for enriching the ARGs. Inclusion of only a limited set of MGEs primers in ARGs qPCR array and also the absence of plasmid specific genes is the major limitation of the MGEs analysis via HT-qPCR which needs to be rectified. More in depth analysis with direct and conclusive experimental designs of validatory nature are needed to determine the route of ARGs enrichment in different studies. Studies for establishing linkages between ARGs and MGEs with certainty have focused on procuring long read sequencing, better ways of sequence assembly and also on source tracking methods based on whole genome sequencing of isolates [11]. For example, MinION can provide a read length of 5000 bp and it has been used to evaluate the genomic regions containing all ARGs, MGEs and virulence genes of Salmonella enterica serovar Typhi H58 [69].

5. Trends of ARGs Diversity in Different Matrices

Only few studies have used and reported diversity indices for depicting the diversity of ARGs [70–72]. However, most of the studies have reported no. of unique genes and class wise percentage of detected ARGs for highlighting the diversity of resistance genes. This may lead to wrong interpretation, because the percentage of detected ARG class is dependent on the primers of the genes present on the qPCR array. For example low detection percentage of sulfonamides in many studies [13,17] is because only few primers (5/384) of sulfonamide resistant genes are present in qPCR array [65]. To get an overall idea, we have calculated the detection frequency for comparing the diversity of ARGs between different environmental matrices. Detection frequency here is defined as no. of unique genes detected in a study divided by total number of reactions analyzed. Total number of reactions are obtained by multiplying number of primers used with the number of samples tested in the
study. The detection frequency of the ARGs in soils of Antarctica was the lowest among others. In past genomic studies have reported that microorganisms can produce a variety of bioactive antimicrobials in pristine environments [73], therefore, the presence of few ARGs in soils of Antarctica is not surprising. In other environmental matrices, the detection frequency is highly variable, which could be due to demographic [74,75] and other anthropogenic factors [76,77]. An interesting and intriguing trend here is the low detection frequency observed for majority of gut microbiome studies (Figure 4). One possible explanation for this trend could be relatively high bacterial diversity in environmental matrices. For example, Fahrenfeld and Bisceglia have explained that sewer lines can contain more diverse bacteria than fecal samples because resistant bacteria from washed off skin, urine, saliva, sputum and so forth, also go into the waste collection systems along with the fecal microbiome [78]. Chances of horizontal gene transfer are more in diverse bacterial populations [79–81]. Hu et al. (2016) has also highlighted, during a comprehensive bacterial mobile resistome study, the existence of potential phylogenetic barrier for horizontal gene transfer of the resistome in bacterial communities [82].

Figure 4. Detection Frequency of different environmental matrices from all included studies.
This explanation was further strengthened when the average normalized gene abundances (ARGs/16S rRNA gene) of different environmental matrices were compared in our study (Figure 5). The animal manure/gut microbiome have lower normalized gene abundances than other matrices. Although the difference was not significant when checked by one way ANOVA analysis but the trend of relatively less ARGs/16S rRNA gene was definitely visible. Information derived by detection frequency may give us an idea about the diversity of the genes at the meta-analysis level but calculation and depiction of ARGs diversity via different diversity indices would be the best option to avoid any confusion and misinterpretation.

Figure 5. Box and whisker plot depicting the log transformed ARG copies per 16S rRNA gene for different environmental samples, (n) represents no. of studies.

Potential in Clinical Laboratories

Although more than 95% of the included studies for ARGs detection and quantification have employed HT-qPCR technology for screening ARGs in various environmental matrices. But this technique also has a great potential and utility in routine clinical laboratories as well. For example a centralized laboratory receiving hundreds of sample per day, where the resistance determination against clinically relevant ARGs is required, can use such an array where limited ARGs can be evaluated against a large sample set. Walker and colleagues have employed a multiplexed nested PCR technique where the 2nd step of the nested PCR was performed on Biomark™ System. An Integrated Fluidic Circuit (Fluidigm) dynamic array capable of analyzing 192 samples against 24 separate PCR assays was used to evaluate 10 antibiotic resistance genes in perianal swab samples collected from patients [83].

6. Implications of Data Analysis and Quality Control

Relative abundance of ARGs has so far been expressed in the form of fold change and/or percentage. There are two main equations which have been utilized for analyzing ARGs relative abundance. One or both have been used in almost all reported studies involving HT-qPCR. The first equation (Equation 1) was developed and introduced in 2008 [84]. Scientists are using this method to analyze the fold change of ARGs with respect to a standard and/or control sample. The other equation (Equation 2) was first reported by Looft and colleagues in 2012 while measuring the antibiotic resistance genes in swine gut microbiome [85]. This equation provides an estimated gene copy number and subsequently relative abundance of ARGs with respect to 16S rRNA gene can be calculated.

\[
\Delta C_T = C_{T(\text{ARG})} - C_{T(16S)}, \quad \Delta \Delta C_T = \Delta C_{T(\text{Target})} - \Delta C_{T(\text{Reference})}
\]

\[
\text{Gene Copy Number} = 10^{(C_{\text{Cut-off}} - C_T)/(10/3)}
\]

HT-qPCR has the capability to perform thousands of reactions at once but questions regarding validity of huge number of reactions can be raised because reproducibility and/or quality check of
individual reactions become almost impossible. Reliable and reproducible background correction and threshold setting should be used to get a reliable $C_T$ value. A $C_T$ cut-off value, representing the limit of detection, has to be selected for Equation (2). $C_T$ cut-off value can influence the estimated gene copy numbers of the ARGs and 16S rRNA gene. Discrepancies in selection of $C_T$ cut-off values are present in various studies, Looft and co-workers in their initial equation has utilized a relatively conservative $C_T$ cut-off value of 26 [85]. Different $C_T$ cut-off values have been used afterwards, $C_T$ cut-off value of as low as 20 [86,87] to as high as 40 [88] have been reported. However in almost half of the studies (25/51) included for the meta-analysis $C_T$ cut-off value of 31 has been used (Figure 6). Certain attempts have been made to address these gaps and disparities, for example, recently Stedtfeld and colleagues have compared two $C_T$ cut-off values (28 vs. 31) while analyzing ARGs in different environmental and clinical samples and found that higher $C_T$ cut-off values can negatively influence the false positive calls. Further $C_T$ cut-off value of 31 was found to overestimate the quantity of the ARGs by 10 folds [65].

![Figure 6](image_url) Figure 6. Bar graph showing Ct cut-off values used for the analysis of HT-qPCR data in the included studies.

HT-qPCR gives very precise results, so even a slight inaccuracy in data analysis can easily result in statistically significant but inaccurate and biased findings. Many studies have reportedly measured 16S rRNA on conventional qPCR platform and ARGs on HT-qPCR platform or with different reaction volumes on same platform [45,72,89]. This can raise concerns because instrumental sensitivity and analytical differences are not often considered. If the reference gene (16S rRNA in this case) is not estimated precisely, subsequent normalization can compromise the results of the whole ARGs qPCR array.

There are very few instances where studies were designed to validate and compare the efficiency and quality of HT-qPCR with other techniques. The analytical sensitivity and specificity of HT-qPCR (Applied Biosystems Open Array) and primer design strategies for reducing the extent of validations for developing qPCR arrays for a greater number of targets have been evaluated [90]. Selective validation of the HT-qPCR results by sequencing amplicons have also been reported [91]. Similarly, validation of specificity for selected primers have been checked by analyzing dilutions of extracted DNA from type strains [71] and also from mock communities [65]. Some studies have also tried to compare ARGs abundance by HT-qPCR and metagenomic analysis [85,92]. Recently Sandberg and colleagues have also run standard curves for each of the 48 target genes included in their study. The standard curves represented a dilution series ranging from $10^8$ to 1 ARG copies per reaction for each gene [48]. Similarly, standard curves for ARGs during another study employing HT-qPCR were also reported [47].

The need of the hour is to put a focus on designing and executing such validation studies for evaluating different environmental matrices. Standard curves are needed to be introduced in HT-qPCR based ARGs investigations. Analysis and gene normalization should follow a standard protocol so that the HT-qPCR results coming from different parts of the world are comparable. If 16S rRNA gene is to be measured on conventional platform, the same sample-assay combination should
also be evaluated on HT-qPCR platform to minimize the error caused by instrumental variations. MIQE guidelines for publishing research related to qPCR should be implemented in true spirit [93]. Research in this area would be extremely useful in formulating precise strategies aimed at minimizing the public health hazards associated with antibiotic resistance. A comprehensive understanding of the HT-qPCR data analysis warrants concerted efforts from molecular biologists and environmental scientists. This will not only enhance the credibility of the HT-qPCR results but will also produce more uniform, comparable and reliable data.

7. Potential of Other Amplification Technologies for ARGs Measurement

The cost of reagents and chemicals is undoubtedly saved in HT-qPCR but the higher cost of HT-qPCR platforms makes it out of the reach of most developing and under developed countries. This is also highlighted by the fact that during the last seven years there is not even a single study from any under developed country (Figure 2). Nucleic acid based point-of-care devices (field-able) can play a critical role in reducing the cost of the platforms [94]. High throughput capacity and automation are possibly some of the key aspects which can be added in these devices. Although a dipstick type of a test is still not possible for ARGs analysis but the need for on-site molecular assessment of ARGs without transporting the samples in a molecular laboratory has already been realized. Scientific community is seeing the exclusion of nucleic acid extraction as a step forward towards on-site ARGs analysis. For example, Kostić and colleagues have reported isothermal amplification of ARGs in 30 min with minimal sample preparation in 384 wells assay cards on a field deployable Gene-Z device [95]. In another isothermal study utilizing MPN-LAMP technique waterborne pathogens were detected in less than 25 min [96]. The use of such techniques in ARGs amplification can bring more simplification in instrumentation further reducing the cost. However there are still many hurdles in the commercialization of such devices possibly due to complexities in primer design and acceptability of amplification methods (except qPCR) in routine diagnostics.

8. Conclusions

Our study has provided an effective assessment of the contributions made by HT-qPCR technology in the field of antimicrobial resistance. The drive to introduce more improvements in the existing technology is likely to continue for the foreseeable future as HT-qPCR finds its way to more laboratories and niches. More sophisticated HT-qPCR platforms and/or field deployable devices with higher levels of automation are expected in near future. In addition to reducing the consumable costs particular attention should be given in reducing the instrument/platform costs and technical complexities for the end user. It is anticipated that in future HT-qPCR based resistance studies will continue to have a focus on analyzing linkages among resistance genes with mobile genetic elements and bacterial community structure to determine the health risk in various environmental samples.

Supplementary Materials: The following are available online.

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References
1. Ciorba, V.; Odone, A.; Veronesi, L.; Pasquaarella, C.; Signorelli, C. Antibiotic resistance as a major public health concern: Epidemiology and economic impact. *Ann. Ig.* 2015, 27, 562–579. [PubMed]
2. Ventola, C.L. The antibiotic resistance crisis: Part 1: Causes and threats. *Pharm. Ther.* 2015, 40, 277–283. [PubMed]
3. Aubertheau, E.; Stalder, T.; Mondamert, L.; Ploy, M.-C.; Dagot, C.; Labanowski, J. Impact of wastewater treatment plant discharge on the contamination of river biofilms by pharmaceuticals and antibiotic resistance. *Sci. Total Environ.* 2017, 579, 1387–1398. [CrossRef] [PubMed]

4. Karkman, A.; Pärnänen, K.; Larsson, D.G.J. Fecal pollution explains antibiotic resistance gene abundances in anthropogenically impacted environments. *bioRxiv* 2018, 341847. [CrossRef]

5. Larsson, D.G.J. Antibiotics in the environment. *Ups. J. Med. Sci.* 2014, 119, 108–112. [CrossRef] [PubMed]

6. Aminov, R.I. Horizontal Gene Exchange in Environmental Microbiota. *Front. Microbiol.* 2011, 2, 158. [CrossRef] [PubMed]

7. Heuer, H.; Smalla, K. Horizontal gene transfer between bacteria. *Environ. Biosafety Res.* 2007, 6, 3–13. [CrossRef]

8. Pruden, A.; Pei, R.; Storteboom, H.; Carlson, K.H. Antibiotic resistance genes as emerging contaminants: Studies in northern Colorado. *Environ. Sci. Technol.* 2006, 40, 7445–7450. [CrossRef]

9. Suzuki, S.; Ogo, M.; Koike, T.; Takada, H.; Newman, B. Sulfonamide and tetracycline resistance genes in total- and culturable-bacterial assemblages in South African aquatic environments. *Front. Microbiol.* 2015, 6, 796. [CrossRef]

10. Jia, S.; He, X.; Bu, Y.; Shi, P.; Miao, Y.; Zhou, H.; Shan, Z.; Zhang, X.-X. Environmental fate of tetracycline resistance genes originating from swine feedlots in river water. *J. Environ. Sci. Heal. Part B* 2014, 49, 624–631. [CrossRef]

11. Waseem, H.; Williams, M.R.; Stedtfeld, T.; Chai, B.; Stedtfeld, R.D.; Cole, J.R.; Tiedje, J.M.; Hashsham, S.A. Virulence factor activity relationships (VFARs): A bioinformatics perspective. *Environ. Sci. Process. Impacts* 2017, 3, 19. [CrossRef] [PubMed]

12. Sun, W.; Qian, X.; Gu, J.; Wang, X.-J.; Duan, M.-L. Mechanism and Effect of Temperature on Variations in Antibiotic Resistance Genes during Anaerobic Digestion of Dairy Manure. *Sci. Rep.* 2016, 6, 30237. [CrossRef] [PubMed]

13. Xu, L.; Ouyang, W.; Qian, Y.; Su, C.; Su, J.; Chen, H. High-throughput profiling of antibiotic resistance genes in drinking water treatment plants and distribution systems. *Environ. Pollut.* 2016, 213, 119–126. [CrossRef] [PubMed]

14. Karkman, A.; Do, T.T.; Walsh, F.; Virta, M.P.J. Antibiotic-Resistance Genes in Waste Water. *Trends Microbiol.* 2018, 26, 220–228. [CrossRef] [PubMed]

15. Chen, Q.-L.; An, X.-L.; Zhu, Y.-G.; Su, J.-Q.; Gillings, M.R.; Ye, Z.-L.; Cui, L. Application of Struvite Alters the Antibiotic Resistome in Soil, Rhizosphere and Phyllosphere. *Environ. Sci. Technol.* 2017, 51, 8149–8157. [CrossRef] [PubMed]

16. Cui, E.-P.; Gao, F.; Liu, Y.; Fan, X.-Y.; Li, Z.-Y.; Du, Z.-J.; Hu, C.; Neal, A.L. Amendment soil with biochar to control antibiotic resistance genes under unconventional water resources irrigation: Proceed with caution. *Environ. Pollut.* 2018, 240, 475–484. [CrossRef] [PubMed]

17. Wang, F.-H.; Qiao, M.; Su, J.-Q.; Chen, Z.; Zhou, X.; Zhu, Y.-G. High Throughput Profiling of Antibiotic Resistance Genes in Urban Park Soils with Reclaimed Water Irrigation. *Environ. Sci. Technol.* 2014, 48, 9079–9085. [CrossRef]

18. de Vries, L.E.; Vallès, Y.; Agersø, Y.; Vaishampayan, P.A.; García-Montaner, A.; Kuehl, J.V.; Christensen, H.; Barlow, M.; Francino, M.P. The Gut as Reservoir of Antibiotic Resistance: Microbial Diversity of Tetracycline Resistance in Mother and Infant. *PLoS ONE* 2011, 6, e21644. [CrossRef]

19. Alicea-Serrano, A.M.; Contreras, M.; Magris, M.; Hidalgo, G.; Dominguez-Bello, M.G. Tetracycline resistance genes acquired at birth. *Arch. Microbiol.* 2013, 195, 447–451. [CrossRef]

20. Van Goethem, M.W.; Pierneef, R.; Bezuidt, O.K.L.; Van De Peer, Y.; Cowan, D.A.; Makhalanyane, T.P. A reservoir of ‘historical’ antibiotic resistance genes in remote pristine Antarctic soils. *Microbiome* 2018, 6, 40. [CrossRef]

21. Waseem, H.; Williams, M.R.; Stedtfeld, R.D.; Hashsham, S.A. Antimicrobial Resistance in the Environment. *Water Environ. Res.* 2017, 89, 921–941. [CrossRef] [PubMed]

22. Case, R.J.; Boucher, Y.; Dahlloff, I.; Holmstrom, C.; Doolittle, W.F.; Kjelleberg, S. Use of 16S rRNA and rpoB Genes as Molecular Markers for Microbial Ecology Studies. *Appl. Environ. Microbiol.* 2007, 73, 278–288. [CrossRef] [PubMed]

23. Klappenbach, J.A.; Saxman, P.R.; Cole, J.R.; Schmidt, T.M. rrndb: The Ribosomal RNA Operon Copy Number Database. *Nucleic Acids Res.* 2001, 29, 181–184. [CrossRef] [PubMed]

24. Babicki, S.; Arndt, D.; Marcu, A.; Liang, Y.; Grant, J.R.; Maciejewski, A.; Wishart, D.S. Heatmapper: Web-enabled heat mapping for all. *Nucleic Acids Res.* 2016, 44, W147–W153. [CrossRef] [PubMed]
25. Hammer, D.A.T.; Ryan, P.D.; Hammer, Ø.; Harper, D.A.T. Past-Paleontological Statistics. Available online: https://www.uv.es/~pardomv/pe/2001_1/past/pastprog/past (accessed on 1 October 2018).

26. Frye, J.G.; Lindsey, R.L.; Rondeau, G.; Porwollik, S.; Long, F.; McClelland, M.; Jackson, C.R.; Englen, M.D.; Meinersmann, R.J.; Berrang, M.E.; et al. Development of a DNA microarray to detect antimicrobial resistance genes identified in the National Center for Biotechnology Information database. Microb. Drug Resist. 2010, 16, 9–19. [CrossRef] [PubMed]

27. Lu, N.; Hu, Y.; Zhu, L.; Yang, X.; Yin, Y.; Lei, F.; Zhu, Y.; Du, Q.; Wang, X.; Meng, Z.; et al. DNA microarray analysis reveals that antibiotic resistance-gene diversity in human gut microbiota is age related. Sci. Rep. 2014, 4, 609. [CrossRef]

28. Perreten, V.; Vorlet-Fawer, L.; Slickers, P.; Ehricht, R.; Kuhnert, P.; Frey, J. Microarray-based detection of 90 antibiotic resistance genes of gram-positive bacteria. J. Clin. Microbiol. 2005, 43, 2291–2302. [CrossRef]

29. Muñir, A.; Waseem, H.; Williams, M.R.; Stedtfeld, R.D.; Gulari, E.; Tiedje, J.M.; Hashsham, S.A. Modeling hybridization kinetics of gene probes in a DNA biochip using FEMLAB. Microarrays 2017, 6, 9. [CrossRef]

30. Jeanty, C.; Longrois, D.; Mertes, P.-M.; Wagner, D.R.; Devaux, Y. An optimized protocol for microarray validation by quantitative PCR using amplified amino allyl labeled RNA. BMC Genomics 2010, 11, 542. [CrossRef]

31. Morey, J.S.; Ryan, J.C.; Van Dolah, F.M. Microarray validation: Factors influencing correlation between oligonucleotide microarrays and real-time PCR. Biol. Proced. Online 2006, 8, 175–193. [CrossRef]

32. Shi, P.; Jia, S.; Zhang, X.X.; Zhang, T.; Cheng, S.; Li, A. Metagenomic insights into chlorination effects on microbial antibiotic resistance in drinking water. Water Res. 2013, 47, 111–120. [CrossRef] [PubMed]

33. Chen, B.; Yang, Y.; Liang, X.; Yu, K.; Zhang, T.; Li, X. Metagenomic Profiles of Antibiotic Resistance Genes (ARGs) between Human Impacted Estuary and Deep Ocean Sediments. Environ. Sci. Technol. 2013, 47, 12753–12760. [CrossRef] [PubMed]

34. Guo, J.; Li, J.; Chen, H.; Bond, P.L.; Yuan, Z. Metagenomic analysis reveals wastewater treatment plants as hotspots of antibiotic resistance genes and mobile genetic elements. Water Res. 2017, 123, 468–478. [CrossRef] [PubMed]

35. Subirats, J.; Sánchez-Melsió, A.; Borrego, C.M.; Balcázar, J.L.; Simonet, P. Metagenomic analysis reveals that bacteriophages are reservoirs of antibiotic resistance genes. Int. J. Antimicrob. Agents 2016, 48, 163–167. [CrossRef] [PubMed]

36. Muziasari, W.I.; Pitkänen, L.K.; Særum, H.; Stedtfeld, R.D.; Tiedje, J.M.; Virta, M. The Resistome of Farmed Fish Feces Contributes to the Enrichment of Antibiotic Resistance Genes in Sediments below Baltic Sea Fish Farms. Front. Microbiol. 2017, 7, 2137. [CrossRef] [PubMed]

37. Muziasari, W.I.; Pärnänen, K.; Johnson, T.A.; Lyra, C.; Karkman, A.; Stedtfeld, R.D.; Tamminen, M.; Tiedje, J.M.; Virta, M. Aquaculture changes the profile of antibiotic resistance and mobile genetic element associated genes in Baltic Sea sediments. FEMS Microbiol. Ecol. 2016, 92, fiw052. [CrossRef] [PubMed]

38. DiCenzo, G.C.; Finan, T.M. The Divided Bacterial Genome: Structure, Function and Evolution. Microbiol. Mol. Biol. Rev. 2017, 81, e00019-17. [CrossRef]

39. Brosius, J.; Palmer, M.L.; Kennedy, P.J.; Noller, H.F. Complete nucleotide sequence of a 16S ribosomal RNA gene from Escherichia coli. Proc. Natl. Acad. Sci. USA. 1978, 75, 4801–4805. [CrossRef]

40. Pinard, R.; de Winter, A.; Sarkis, G.J.; Gerstein, M.B.; Tartaro, K.R.; Plant, R.N.; Egholm, M.; Rothberg, J.M.; Leamon, J.H. Assessment of whole genome amplification-induced bias through high-throughput, massively parallel whole genome sequencing. BMC Genomics 2006, 7, 216. [CrossRef]

41. Giardina, E.; Pietrangeli, I.; Martone, C.; Zampatti, S.; Marsala, P.; Gabriele, L.; Ricci, O.; Solla, G.; Asili, P.; Arcudi, G.; et al. Whole genome amplification and real-time PCR in forensic casework. BMC Genomics 2009, 10, 159. [CrossRef]

42. Sipos, R.; Székely, A.J.; Palatinzsky, M.; Révész, S.; Márialigeti, K.; Nikolaus, M. Effect of primer mismatch, annealing temperature and PCR cycle number on 16s rRNA gene-targetting bacterial community analysis. FEMS Microbiol. Ecol. 2007, 60, 341–350. [CrossRef]

43. Malhotra, K.; Foltz, L.; Mahoney, W.C.; Schueler, P.A. Interaction and effect of annealing temperature on primers used in differential display RT-PCR. Nucleic Acids Res. 1998, 26, 854–856. [CrossRef] [PubMed]

44. Kuang, J.; Yan, X.; Genders, A.J.; Granata, C.; Bishop, D.J. An overview of technical considerations when using quantitative real-time PCR analysis of gene expression in human exercise research. PLoS ONE 2018, 13, e0196438. [CrossRef] [PubMed]
45. Hu, H.-W.; Wang, J.-T.; Li, J.; Li, J.-J.; Ma, Y.-B.; Chen, D.; He, J.-Z. Field-based evidence for copper contamination induced changes of antibiotic resistance in agricultural soils. *Environ. Microbiol.* 2016, 18, 3896–3909. [CrossRef] [PubMed]

46. Hu, H.-W.; Wang, J.-T.; Li, J.; Shi, X.-Z.; Ma, Y.-B.; Chen, D.; He, J.-Z. Long-Term Nickel Contamination Increases the Occurrence of Antibiotic Resistance Genes in Agricultural Soils. *Environ. Sci. Technol.* 2017, 51, 790–800. [CrossRef] [PubMed]

47. Ahmed, W.; Zhang, Q.; Lobos, A.; Senkeil, J.; Sadowsky, M.J.; Harwood, V.J.; Saedi, N.; Marinoni, O.; Ishii, S. Precipitation influences pathogenic bacteria and antibiotic resistance gene abundance in storm drain outfalls in coastal sub-tropical waters. *Environ. Int.* 2018, 116, 308–318. [CrossRef] [PubMed]

48. Sandberg, K.D.; Ishii, S.; LaPara, T.M. A Microfluidic Quantitative Polymerase Chain Reaction Method for the Simultaneous Analysis of Dozens of Antibiotic Resistance and Heavy Metal Resistance Genes. *Environ. Sci. Technol. Lett.* 2018, 5, 20–25. [CrossRef]

49. Xiong, W.; Sun, Y.; Ding, X.; Wang, M.; Zeng, Z. Selective pressure of antibiotics on ARGs and bacterial communities in manure-polluted freshwater-sediment microcosms. *Front. Microbiol.* 2015, 6, 194. [CrossRef]

50. Tian, Z.; Zhang, Y.; Yu, B.; Yang, M. Changes of resistome, mobilome and potential hosts of antibiotic resistance genes during the transformation of anaerobic digestion from mesophilic to thermophilic. *Water Res.* 2016, 98, 261–269. [CrossRef]

51. Von Wintersdorff, C.J.H.; Penders, J.; van Niekerk, J.M.; Mills, N.D.; Majumder, S.; van Alphen, L.B.; Savelkoul, P.H.M.; Wolfs, P.F.G. Dissemination of Antimicrobial Resistance in Microbial Ecosystems through Horizontal Gene Transfer. *Front. Microbiol.* 2016, 7, 173. [CrossRef]

52. De Faria, A.C.S.; de Godoy, I.; Sanches, A.A.A.; Iglesias, G.A.; Candido, S.L.; da Paz, R.C.R.; Nakazato, L.; Dutra, V.; de Faria, A.C.S.; de Godoy, I.; et al. Detection of resistance genes and evaluation of water quality at zoo lakes in Brazil. *Ciência Rural* 2016, 46, 860–866. [CrossRef]

53. Zheng, J.; Gao, R.; Wei, Y.; Chen, T.; Fan, J.; Zhou, Z.; Makimilua, T.B.; Jiao, Y.; Chen, H. High-throughput profiling and analysis of antibiotic resistance genes in East Tiaoxi River, China. *Environ. Pollut.* 2017, 230, 648–654. [CrossRef] [PubMed]

54. Xiang, Q.; Chen, Q.-L.; Zhu, D.; An, X.-L.; Yang, X.-R.; Su, J.-Q.; Qiao, M.; Zhu, Y.-G. Spatial and temporal distribution of antibiotic resistomes in a peri-urban area is associated significantly with anthropogenic activities. *Environ. Pollut.* 2018, 235, 525–533. [CrossRef] [PubMed]

55. Chang, H.-H.; Cohen, T.; Grad, Y.H.; Hanage, W.P.; O’Brien, T.F.; Lipsitch, M. Origin and proliferation of multiple-drug resistance in bacterial pathogens. *Microbiol. Mol. Biol. Rev.* 2015, 79, 101–116. [CrossRef] [PubMed]

56. Mc Mahon, M.A.S.; Blair, I.S.; Moore, J.E.; Mc Dowell, D.A. The rate of horizontal transmission of antibiotic resistance plasmids is increased in food preservation-stressed bacteria. *J. Appl. Microbiol.* 2007, 103, 1883–1888. [CrossRef] [PubMed]

57. Jacquiod, S.; Brejnrod, A.; Morberg, S.M.; Abu Al-Soud, W.; Sørensen, S.J.; Riber, L. Deciphering conjugative plasmid permissiveness in wastewater microbiomes. *Mol. Ecol.* 2017, 26, 3556–3571. [CrossRef] [PubMed]

58. Stokes, H.W.; Gillings, M.R. Gene flow, mobile genetic elements and the recruitment of antibiotic resistance genes during the transformation of anaerobic digestion from mesophilic to thermophilic. *Water Res.* 2016, 98, 261–269. [CrossRef]

59. Loot, C.; Nivina, A.; Cory, J.; Escudero, J.A.; Ducos-Galand, M.; Bikhad, D.; Rocha, E.P.; Mazel, D. Differences in Integron Cassette Excision Dynamics Shape a Trade-Off between Evolvability and Genetic Capacitance. *MBio* 2017, 8, e02296-16. [CrossRef] [PubMed]

60. Gillings, M.R.; Gaze, W.H.; Pruden, A.; Smalla, K.; Tiedje, J.M.; Zhu, Y.G. Using the class 1 integron-integrase gene as a proxy for anthropogenic pollution. *ISME J.* 2015, 9, 1269–1279. [CrossRef]

61. Stedtfeld, R.D.; Stedtfeld, T.M.; Wassem, H.; Fitschen-Brown, M.; Guo, X.; Chai, B.; Williams, M.R.; Shook, T.; Logan, A.; Graham, A.; et al. Isothermal assay targeting class 1 integrase gene for environmental surveillance of antibiotic resistance markers. *J. Environ. Manage.* 2017, 198, 213–220. [CrossRef]

62. Pruden, A.; Arabi, M.; Storteboom, H.N. Correlation Between Upstream Human Activities and Riverine Antibiotic Resistance Genes. *Environ. Sci. Technol.* 2012, 46, 11541–11549. [CrossRef] [PubMed]

63. Stedtfeld, R.D.; Williams, M.R.; Fakher, U.; Johnson, T.A.; Stedtfeld, T.M.; Wang, F.; Khalife, W.T.; Hughes, M.; Etchebarne, B.E.; Tiedje, J.M.; et al. Antimicrobial resistance dashboard application for mapping environmental occurrence and resistant pathogens. *FEMS Microbiol. Ecol.* 2016, 92, fiw020. [CrossRef] [PubMed]
64. Le, T.-H.; Ng, C.; Chen, H.; Yi, X.Z.; Koh, T.H.; Barkham, T.M.S.; Zhou, Z.; Gin, K.Y.-H. Occurrences and Characterization of Antibiotic Resistant Bacteria and Genetic Determinants of Hospital Wastewater in a Tropical Country. Antimicrob. Agents Chemother. 2016, 60, 7449–7456. [CrossRef] [PubMed]
65. Stedtfeld, R.D.; Guo, X.; Stedtfeld, T.M.; Sheng, H.; Williams, M.R.; Hauschild, K.; Gunturu, S.; Tift, L.; Wang, F.; Howe, A.; et al. Primer set 2.0 for highly parallel qPCR array targeting antibiotic resistance genes and mobile genetic elements. FEMS Microbiol. Ecol. 2018, 94, fiy130. [CrossRef] [PubMed]
66. Chakravorty, S.; Helb, D.; Burday, M.; Connell, N.; Alland, D. A detailed analysis of 16S ribosomal RNA gene segments for the diagnosis of pathogenic bacteria. J. Microbiol. Methods 2007, 69, 330–339. [CrossRef] [PubMed]
67. Guo, F.; Ju, F.; Cai, L.; Zhang, T. Taxonomic Precision of Different Hypervariable Regions of 16S rRNA Gene and Annotation Methods for Functional Bacterial Groups in Biological Wastewater Treatment. PLoS ONE 2013, 8, e76185. [CrossRef]
68. Waldron, L.S.; Gillings, M.R. Screening Foodstuffs for Class 1 Integrons and Gene Cassettes. J. Vis. Exp. 2015, 100, e52889. [CrossRef] [PubMed]
69. Ashton, P.M.; Nair, S.; Dallman, T.; Rubin, S.; Rabsch, W.; Mwaigwisya, S.; Wain, J.; O'Grady, J. MinION nanopore sequencing identifies the position and structure of a bacterial antibiotic resistance island. Nat. Biotechnol. 2015, 33, 296–300. [CrossRef]
70. Chen, Q.; An, X.; Li, H.; Su, J.; Ma, Y.; Zhu, Y.-G. Long-term field application of sewage sludge increases the abundance of antibiotic resistance genes in soil. Environ. Int. 2016, 92–93, 1–10. [CrossRef]
71. Zhu, Y.G.; Johnson, T.A.; Su, J.Q.; Qiao, M.; Guo, G.X.; Stedtfeld, R.D. Diverse and abundant antibiotic resistance genes in Chinese swine farms. Proc. Natl. Acad. Sci. USA 2013, 110, 1–6. [CrossRef]
72. Ouyang, W.-Y.; Huang, F.-Y.; Zhao, Y.; Li, H.; Su, J.-Q. Increased levels of antibiotic resistance in urban stream of Jiulongjiang River, China. Appl. Microbiol. Biotechnol. 2015, 99, 5697–5707. [CrossRef] [PubMed]
73. Bernier, S.P.; Surette, M.G. Concentration-dependent activity in natural environments. Front. Microbiol. 2013, 4, 20. [CrossRef]
74. Singh, A.K.; Das, S.; Singh, S.; Gajamer, V.R.; Pradhan, N.; Lepcha, Y.D.; Tiwari, H.K. Prevalence of antibiotic resistance in commensal Escherichia coli among the children in rural hill communities of Northeast India. PLoS ONE 2018, 13, e0199179. [CrossRef] [PubMed]
75. Hoban, D.; Baquero, F.; Reed, V.; Felmingham, D. Demographic analysis of antimicrobial resistance among Streptococcus pneumoniae: Worldwide results from PROTEKT 1999–2000. Int. J. Infect. Dis. 2005, 9, 262–273. [CrossRef] [PubMed]
76. Castro-Sánchez, E.; Moore, L.S.P.; Husson, F.; Holmes, A.H. What are the factors driving antimicrobial resistance? Perspectives from a public event in London, England. BMC Infect. Dis. 2016, 16, 465. [CrossRef] [PubMed]
77. Allcock, S.; Young, E.H.; Holmes, M.; Gurudasani, D.; Dougan, G.; Sandhu, M.S.; Solomon, L.; Török, M.E. Antimicrobial resistance in human populations: Challenges and opportunities. Glob. Heal. Epidemiol. Genomics 2017, 2, e4. [CrossRef] [PubMed]
78. Fahrenfeld, N.; Biseglia, K.J. Emerging investigators series: Sewer surveillance for monitoring antibiotic use and prevalence of antibiotic resistance: Urban sewer epidemiology. Environ. Sci. Water Res. Technol. 2016, 2, 788–799. [CrossRef]
79. Townsend, J.P.; Behn, T.; Nielsen, K.M. Assessing the probability of detection of horizontal gene transfer events in bacterial populations. Front. Microbiol. 2012, 3, 27. [CrossRef]
80. Dutta, C.; Pan, A. Horizontal gene transfer and bacterial diversity. J. Biosci. 2002, 27, 27–33. [CrossRef]
81. Liu, L.; Chen, X.; Skogerbø, G.; Zhang, F.; Chen, R.; He, S.; Huang, D.-W. The human microbiome: A hot spot of microbial horizontal gene transfer. Genomics 2012, 100, 265–270. [CrossRef]
82. Hu, Y.; Yang, X.; Li, J.; Lv, N.; Liu, F.; Wu, J.; Lin, L.Y.C.; Wu, N.; Weimer, B.C.; Gao, G.F.; et al. The Bacterial Mobile Resistome Transfer Network Connecting the Animal and Human Microbiomes. Appl. Environ. Microbiol. 2016, 82, 6672–6681. [CrossRef] [PubMed]
83. Walker, G.T.; Rockweiler, T.J.; Kersey, R.K.; Frye, K.L.; Mitchner, S.R.; Toal, D.R.; Quan, J. Analytical Performance of Multiplexed Screening Test for 10 Antibiotic Resistance Genes from Perianal Swab Samples. Clin. Chem. 2016, 62, 353–359. [CrossRef] [PubMed]
84. Schmittgen, T.D.; Livak, K.J. Analyzing real-time PCR data by the comparative C(T) method. Nat. Protoc. 2008, 3, 1101–1108. [CrossRef] [PubMed]
85. Looft, T.; Johnson, T.A.; Allen, H.K.; Bayles, D.O.; Alt, D.P.; Stedtfeld, R.D.; Sul, W.J.; Stedtfeld, T.M.; Chai, B.; Cole, J.R.; et al. In-feed antibiotic effects on the swine intestinal microbiome. *Proc. Natl. Acad. Sci. USA* **2012**, *109*, 1691–1696. [CrossRef] [PubMed]

86. Buelow, E.; Bayjanov, J.R.; Majoor, E.; Willems, R.J.; Bonten, M.J.; Schmitt, H.; van Schaik, W. Limited influence of hospital wastewater on the microbiome and resistome of wastewater in a community sewerage system. *FEMS Microbiol. Ecol.* **2018**, *94*, fiy087. [CrossRef] [PubMed]

87. Buelow, E.; Bello González, T.D.J.; Fuentes, S.; de Steenhuijsen Piers, W.A.A.; Lahti, L.; Bayjanov, J.R.; Majoor, E.A.M.; Braat, J.C.; van Mourik, M.S.M.; Oostdijk, E.A.N.; et al. Comparative gut microbiota and resistome profiling of intensive care patients receiving selective digestive tract decontamination and healthy subjects. *Microbiome* **2017**, *5*, 88. [CrossRef] [PubMed]

88. Wang, B.; Li, G.; Cai, C.; Zhang, J.; Liu, H. Assessing the safety of thermally processed penicillin mycelial dreg following the soil application: Organic matter’s maturation and antibiotic resistance genes. *Sci. Total Environ.* **2018**, *636*, 1463–1469. [CrossRef]

89. Kang, W.; Zhang, Y.-J.; Shi, X.; He, J.-Z.; Hu, H.-W. Short-term copper exposure as a selection pressure for antibiotic resistance and metal resistance in an agricultural soil. *Environ. Sci. Pollut. Res.* **2018**, *25*, 29314–29324. [CrossRef]

90. Stedtfeld, R.D.; Baushke, S.W.; Tournoulle, D.M.; Miller, S.M.; Stedtfeld, T.M.; Gulari, E.; Tiedje, J.M.; Hashsham, S.A. Development and Experimental Validation of a Predictive Threshold Cycle Equation for Quantification of Virulence and Marker Genes by High-Throughput Nanoliter-Volume PCR on the OpenArray Platform. *Appl. Environ. Microbiol.* **2008**, *74*, 3831–3838. [CrossRef]

91. Johnson, T.A.; Stedtfeld, R.D.; Wang, Q.; Cole, J.R.; Hashsham, S.A.; Looft, T.; Zhu, Y.-G.; Tiedje, J.M. Clusters of Antibiotic Resistance Genes Enriched Together Stay Together in Swine Agriculture. *MBio* **2016**, *7*, e02214–e02215. [CrossRef]

92. Stedtfeld, R.D.; Stedtfeld, T.M.; Fader, K.A.; Williams, M.R.; Bhaduri, P.; Quensen, J.; Zacharewski, T.R.; Tiedje, J.M.; Hashsham, S.A. TCDD influences reservoir of antibiotic resistance genes in murine gut microbiome. *FEMS Microbiol. Ecol.* **2017**, *93*, fx058. [CrossRef] [PubMed]

93. Bustin, S.A.; Benes, V.; Garson, J.A.; Hellemans, J.; Huggett, J.; Kubista, M.; Mueller, R.; Nolan, T.; Pfaffl, M.W.; Shipley, G.L.; et al. The MIQE Guidelines: Minimum Information for Publication of Quantitative Real-Time PCR Experiments. *Clin. Chem.* **2009**, *55*, 611–622. [CrossRef] [PubMed]

94. Williams, M.R.; Stedtfeld, R.D.; Waseem, H.; Stedtfeld, T.; Upham, B.; Khalife, W.; Etchebarne, B.; Hughes, M.; Tiedje, J.M.; Hashsham, S.A. Implications of direct amplification for measuring antimicrobial resistance using point-of-care devices. *Anal. Methods* **2017**, *9*, 1229–1241. [CrossRef] [PubMed]

95. Kostić, T.; Ellis, M.; Williams, M.R.; Stedtfeld, T.M.; Kaneene, J.B.; Stedtfeld, R.D.; Hashsham, S.A. Thirty-minute screening of antibiotic resistance genes in bacterial isolates with minimal sample preparation in static self-dispensing 64 and 384 assay cards. *Appl. Microbiol. Biotechnol.* **2015**, *99*, 7711–7722. [CrossRef] [PubMed]

96. Ahmad, F.; Stedtfeld, R.D.; Waseem, H.; Williams, M.R.; Cupples, A.M.; Tiedje, J.M.; Hashsham, S.A. Most probable number—Loop mediated isothermal amplification (MPN-LAMP) for quantifying waterborne pathogens in <25 min. *J. Microbiol. Methods* **2017**, *33*, 27–33. [CrossRef] [PubMed]

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