Heavy metals in soils and forage grasses irrigated with Vieira River water, Montes Claros, Brazil, contaminated with sewage wastewater

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ABSTRACT

There is great concern with soil and plant contamination by heavy metals due to the use of polluted water in agricultural irrigation. In this study, areas irrigated with Vieira River water were evaluated as to contamination by As, Cr, Cu, Ni, Pb and Zn. The Vieira River receives effluent from Montes Claros city, state of Minas Gerais, Brazil. To do so, two irrigated areas were selected, one upstream and one downstream of the Montes Claros city. Wastewater discharge increased the concentration of As and Ni in the water of Vieira River, and consequently, of As, Cr, Cu, Ni, Pb and Zn in the soil and of As and Zn in forage grasses. However, the content of heavy metals in the soil did not exceed the internationally recommended limits. Pollution load index (PLI) and contamination factor (CF) indicated the existence of pollution and moderate contamination in downstream soils of the city of Montes Claros. Potential ecological risk index (RI) and ecological risk factor (Er) indicated a low ecological risk, but these indicators were higher in downstream soils of Montes Claros. Arsenic (As) was the only heavy metal that featured a transfer factor (TF) higher than the widespread values found in literature and positive geoaccumulation index (Igeo), indicative of anthropogenic pollution.

Keywords: ecological risk, geoaccumulation index, water contaminated.

Metais pesados em solos e plantas forrageiras irrigadas com água do Rio Vieira, Montes Claros, Brasil, contaminadas com efluentes de esgoto

RESUMO

A contaminação de solo e plantas por metais pesados devido ao uso de águas poluídas na irrigação agrícola é motivo de grande preocupação. No presente estudo, áreas irrigadas com água do rio Vieira foram avaliadas quanto a contaminação por As, Cr, Cu, Ni, Pb e Zn. Para isso foram selecionadas duas áreas, uma a montante e outra a jusante da cidade de Montes Claros. A descarga de água residuária aumentou a concentração de As e Ni na água do rio...
Vieira; and consequently, As, Cr, Cu, Ni, Pb, and Zn in soil; and As and Zn in forage grasses. However, the heavy metal concentrations in soil did not exceed the international recommended limits. The pollution load index (PLI) and the contamination factor (CF) indicated moderate pollution and contamination in soils downstream of Montes Claros. The As was the only heavy metal that presented a transmission factor (TF) greater than the generalized values found in the literature and the geoaccumulation index ($I_{geo}$) positive, indicative of anthropogenic pollution.

**Palavras-chave:** água contaminada, índice de geoacumulação, risco ecológico.

### 1. INTRODUCTION

The dispersal of heavy metals in irrigated agricultural areas is growing, which results in food contamination that can be harmful for animals and humans. Heavy metals have low solubility and are not degradable in water, which contributes to their accumulation in soil, and therefore in plants cultivated on these irrigated areas with contaminated water. Heavy metals’ transfer from soils to plants is one of the main ways of humans exposure through the food chain (Chopra and Pathak, 2015).

Until 2010, Vieira River was the main river receiving untreated sewage produced in the city of Montes Claros, state of Minas Gerais, Brazil. Currently, the river still receives city sewage; however, there is a prior treatment process, carried out in the wastewater treatment plant (WWTP) of Montes Claros. In the margins of Vieira River, there are farms that use its water for the irrigation of pastures; nevertheless, there are no studies on heavy metal contamination of soil and plants irrigated with the river water.

The intake of heavy metals through the food chain has been widely disseminated throughout the world (Chary et al., 2008; Chopra and Pathak, 2015). Heavy metals are non-biodegradable and persistent compounds, which, because of these characteristics, accumulate in vital organs of the human body, such as kidneys, bones, and liver, and are associated with several serious health disorders, such as diarrhea, stomatitis, tremor, ataxia, paralysis, convulsion, depression, and pneumonia (McCluggage, 1991). The nature of the effects of these disorders can be toxic, neurotoxic, carcinogenic, mutagenic, or teratogenic (EU, 2002).

This study assessed the concentration of As, Cr, Cu, Ni, Pb, and Zn in water, soil, and forage grasses in areas irrigated with Vieira River water.

### 2. MATERIALS AND METHODS

#### 2.1. Study area

The study area is located in the city of Montes Claros, Brazil, in the North of the state of Minas Gerais. The climate of the region is classified, according to Köppen, as Aw. Vieira River is the main receiver of wastewater of domestic and industrial sewage of Montes Claros. Vieira River rises in the south of Montes Claros and flows through residential and industrial areas before joining Verde Grande River, one of the main affluents of the middle region of the São Francisco River Basin. The length of the main course of Vieira River and the area of its basin is approximately 42 km and 580 km$^2$, respectively. The Wastewater Treatment Plant (WWTP) is located in Northern Montes Claros (Figure 1). Several industry plants are located west of the WWTP (Figure 1), including textile, biofuels, food, pharmaceutical, and mining industries. About 1 km east of the WWTP, there is another textile factory (Figure 1).
The unpleasant odor and presence of solid residues and white foam in waters of Vieira River are easily observed in stretches passing through Montes Claros and downstream of WWTP.

2.2. Sample collection

To evaluate levels of contamination by heavy metals in areas irrigated with Vieira River water, samples of water, soil, and plants were collected at two farms (sites F1 and F2 in Figure 1) in July 2015. The F1 farm is located upstream of the city of Montes Claros and uses Vieira River water for irrigation of forage grasses (Pennisetum purpureum), and the irrigation system used is the sprinkler. The F2 farm is located downstream of the city of Montes Claros and diverts part of the Vieira River water to an area of flood irrigation, which cultivates forage grasses (Brachiaria mutica) for feeding the cattle of milk producers.

Samples of unfiltered water of the river were collected approximately 10 m from the margin, where the water used in irrigation is collected. The water samples were stored in polypropylene bottles that had been previously washed with HNO₃ (10%). To preserve the water samples, about 1.5 mL of HNO₃ (65%) was added to reduce their pH to values lower than 2. Sampling bottles were kept in coolers with ice during their transportation to the laboratory, where they were stored in a refrigerator at temperatures between 0 and 4°C until the quantification of heavy metals (APHA et al., 1998).

At each farm (F1 and F2), eight sites were selected for the collection of samples of soil and forage grasses (Figure 1). Each sample was composed of three subsamples collected in an area of 1 m². Samples were placed in plastic bags and sealed for transportation and storage (Gebrekidan et al., 2013).

Soil samples were collected in the 0-20 cm layer. In the laboratory, samples were dried in an oven at 60°C for 72 hours, homogenized in an agate mortar and pistil, and sieved in a Nylon sieve of 0.3 mm.

In the samplings of plant material, the plants were cut to 4 cm from the ground. In the laboratory, samples were washed with tap water and distilled water for removing particulate material deposited on the surface of the plant. Then, all samples were dried in an oven at 75°C.
until constant weight, grinded, and sieved (Nylon sieve of 0.3 mm).

The sieved soil and plant samples were stored in Falcon tubes for further analysis of heavy metal content.

2.3. Chemical analysis

The stock standard solution of 1000 mg L\(^{-1}\) for atomic absorption spectroscopy of Sigma Aldrich (Germany), diluted in different concentrations with ultrapure water (Milli-Q, Millipore), was used for the preparation of analytical curves. Cr, Cu, Ni, Pb, and Zn elements were measured in flame atomization mode, composed of Ar-acetylene (acetylene 2.8 AA), and the As element was determined by a hydride generator system using argon (99.999% purity) and Ar-acetylene (acetylene 2.8 AA) (APHA \textit{et al.}, 1998). All analyses were performed in triplicate.

For total metal analysis, 100 mL of the water sample was mixed with 10 mL of HNO\(_3\) (65%, PA) and taken to a hot plate, in which it was evaporated to a volume from 10 to 20 mL (APHA \textit{et al.}, 1998). Digested samples were then filtered through a quantitative filter paper (Unifil, C42, blue stripe) and transferred to Falcon tubes. The volume of each sample was adjusted to 25 mL using ultrapure water (Milli-Q, Millipore). Concentrations of As, Cr, Cu, Ni, Pb, and Zn were determined in an atomic absorption spectrophotometer (Varian, AA 240 FS model).

For soil digestion, 0.50 g of each sample was mixed to 9 mL of HNO\(_3\) (65%, PA) and 3 mL of HCl (37%) and microwave digested (Mars 6, CEM) according to the USEPA 3051A method (USEPA, 1998). In the case of forage grasses, 0.01 g of each sample was mixed with 10 mL of HNO\(_3\) (65%, PA) and digested in a similar way to the soil samples.

After the digestion process, soil and plant samples were then filtered through a quantitative filter paper (Unifil, C42, blue stripe) and diluted with ultrapure water (Milli-Q, Millipore) until it reached a volume of 25 mL. Concentrations of As, Cr, Cu, Ni, Pb, and Zn were determined in an atomic absorption spectrophotometer (Varian, AA 240 FS model).

2.4. Indicators and Statistical analysis

To understand the impact of heavy metals present in the soil, the following were calculated: contamination factor (CF) (Equation 1); pollution load index (PLI) (Equation 2); geoaccumulation index (\(I_{geo}\)) (Equation 3); and potential ecological risk index (RI) (Equation 4). Assessment of contamination level by heavy metals using these factors has been successfully employed in several studies (Varol, 2011; Wang \textit{et al.}, 2018).

\[
CF = \frac{c_{\text{heavy metal}}}{c_{\text{background}}} \quad (1)
\]

CF is the ratio between each metal content in soil samples and the background value. The adopted background values were 0.23 mg kg\(^{-1}\) for As (Kabata-Pendias and Pendias, 1999); 35.00; 25.00; 20.00; and 20.00 mg kg\(^{-1}\) for Cr, Cu, Ni, and Pb, respectively (Taylor and McLennan, 1995); and 49.50 mg kg\(^{-1}\) for Zn (Turekian and Wedepohl, 1961).

Interpretation of CF values was suggested by Hakanson (1980), according to which CF < 1 indicates low contamination; 1 < CF < 3 indicates moderate contamination; 3 < CF < 6 indicates considerable contamination; CF > 6 indicates very high contamination.

\[
PLI = (CF_1 \times CF_2 \times CF_3 \times \ldots \times CF_n)^{\frac{1}{n}} \quad (2)
\]

Where CF is the contamination factor, and \(n\) is the number of metal species. PLI presents a general overview of pollution in each sampling site and also shows the contribution of each
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metal to pollution. PLI < 1 indicates that there is no pollution by heavy metals in the site; however, PLI > 1 indicates that there is pollution (Varol, 2011).

$I_{geo}$ determines metal contamination in soils, comparing current contents with background values (Equation 3).

$$I_{geo} = \left[ \frac{C_n}{1.5B_n} \right]$$ (3)

Where $C_n$ is the $n$ metal concentration in soil samples, and $B_n$ is the background value of the $n$ element. The factor 1.5 is introduced in the equation to minimize possible variation in the background value, which can be attributed to lithospheric effects. The seven categories of the geoaccumulation index are presented as follows: $I_{geo} < 0$, practically uncontaminated; $0 \leq I_{geo} < 1$, uncontaminated to moderately polluted; $1 \leq I_{geo} < 2$, moderately polluted; $2 \leq I_{geo} < 3$, moderately to heavily polluted; $3 \leq I_{geo} < 4$, heavily polluted; $4 \leq I_{geo} < 5$, heavily to extremely polluted; and $I_{geo} > 5$, extremely polluted.

$RI$ is a factor used to evaluate the potential ecological risk of heavy metals in soils and was initially proposed by Hakanson (1980) (Equation 4).

$$RI = \sum_{i=1}^{n} E_r = \sum_{i=1}^{n} T_r \times CF$$ (4)

Where $E_r$ is the ecological risk factor of a certain heavy metal and $T_r$ is the toxicity factor of a single metal. Toxicity factors of As, Cr, Cu, Ni, Pb, and Zn were 10, 2, 5, 5, 5, and 1, respectively (Hakanson, 1980). Criteria for the $E_r$ classification are: $E_r < 40$, low potential ecological risk; $40 \leq E_r < 80$, moderate potential ecological risk; $80 \leq E_r < 160$, considerable potential ecological risk; $160 \leq E_r < 320$, high potential ecological risk; and $E_r \geq 320$, very high potential ecological risk (Hakanson, 1980). The RI classification introduced by Hakanson (1980) was based on eight parameters (As, Cd, Cr, Cu, Hg, Pb, Zn, and PCBs). In this study, all these parameters were not evaluated, therefore, the RI classification should be adjusted according to the number of parameters and the ratio of toxicity coefficient of each parameter (Wang et al., 2018). The adjusted RI classification is: $RI < 110$, low ecological risk; $110 \leq RI < 220$, moderate ecological risk; $220 \leq RI < 440$, considerable ecological risk; and $RI \geq 440$, very high ecological risk.

For understanding the bioavailability of a heavy metal, was calculated the transfer factor (TF), which in this study was determined by the relationship between the metal content in the shoot of the plants and in the soil (Equation 5) (Gebrekidan et al., 2013).

$$TF = \frac{\text{Metal content in the shoot of the plant}}{\text{Metal content in the soil}}$$ (5)

Analysis of variance (ANOVA) and F-test were conducted at 5% error probability level to verify differences in contamination by heavy metals between upstream (F1) and downstream (F2) of the city of Montes Claros. All statistical analyses were carried out using the R-plus® 3.4.2. software (R Development Core Team, 2017).

3. RESULTS AND DISCUSSION

3.1. Heavy metal contamination in water of Vieira River used in irrigation

The concentrations of heavy metals As and Ni were significantly higher (p < 0.01) in the water of Vieira River at the farm margins downstream of the city of Montes Claros (F2), when compared with the water of the river at the farm margins upstream (F1). On the other hand, Cu featured the higher concentration (p < 0.01) in the water of Vieira River at the F1 margins. None of the concentration values of heavy metals found in the two farms (F1 and F2) were higher than the international limits recommended for water use in irrigation (USEPA, 2012;
WHO, 2006). However, Cu and Ni exceeded the limits recommended by the Brazilian legislation for the use of water in irrigation (CONAMA, 2009), being Cu in the two farms (F1 and F2) and Ni only in F2 (Table 1). It is important to emphasize that the water can be inappropriate for irrigation and superficial flooding depending on the use and occupation of the soil, once this irrigation method can favor the direct contact of animals and humans with the contaminated water.

Table 1. Concentration of heavy metals (µg L\(^{-1}\)) in water samples of Vieira River for irrigation of F1 and F2 farms, and maximum permitted concentration (µg L\(^{-1}\)) in water (n = 3).

| Metals (µg L\(^{-1}\)) | F1 Farm | F2 Farm | CONAMA (2005) | USEPA (2012) | WHO (2006) |
|-------------------------|---------|---------|---------------|--------------|-------------|
|                         | Mean    | SD      | Mean          | SD           |             |
| As (µg L\(^{-1}\))     | 0.4**   | 0.09    | 1.7           | 0.01         | 33          | 100         | 100         |
| Cr (µg L\(^{-1}\))     | 0.0     | 0.00    | 0.0           | 0.00         | 50          | 100         | 100         |
| Cu (µg L\(^{-1}\))     | 82.6**  | 0.74    | 19.4          | 2.61         | 13          | 200         | 200         |
| Ni (µg L\(^{-1}\))     | 0.0**   | 0.00    | 52.3          | 1.51         | 25          | 200         | 200         |
| Pb (µg L\(^{-1}\))     | 0.0     | 0.00    | 0.0           | 0.00         | 33          | 5000        | 5000        |
| Zn (µg L\(^{-1}\))     | 189.1   | 5.94    | 197.6         | 23.25        | 5000        | 2000        | 2000        |

SD – Standard deviation; and ** (p < 0.01) indicates statistical difference (ANOVA; F-test) between mean values of metals in F1 and F2.

3.2. Heavy metal contamination in soil irrigated with water of Vieira River

Contents of As, Cr, Cu, Ni, Pb, and Zn in samples of soil irrigated with water of Vieira River in F2 were significantly higher (p < 0.05) than in F1 (Table 2), which indicates anthropogenic pollution of Vieira River waters, possibly due to the discharge of domestic and industrial wastewater generated in Montes Claros. The highest concentration of As and Ni in water samples collected at F2 (Table 1) also reinforces the indications of anthropogenic pollution of Vieira River waters.

Table 2. Content of heavy metals (mg kg\(^{-1}\)), contamination factor (CF), and pollution load index (PLI) in soil irrigated with Vieira River water in two farms (F1 and F2) (n = 8).

| Metals (mg kg\(^{-1}\)) | F1 Farm | F2 Farm | Maximum allowed |
|-------------------------|---------|---------|-----------------|
|                         | Mean    | SD      | CF             | Mean | SD      | CF     |
| As (mg kg\(^{-1}\))     | 0.3     | 0.05    | 1.40           | 0.6**| 0.12    | 2.50   | 20.0 | 35.0 |
| Cr (mg kg\(^{-1}\))     | 30.5    | 5.33    | 0.87           | 42.5**| 5.27    | 1.21   | 100.0| 150.0|
| Cu (mg kg\(^{-1}\))     | 15.9    | 4.01    | 0.64           | 20.0* | 1.73    | 0.80   | 100.0| 200.0|
| Ni (mg kg\(^{-1}\))     | 5.5     | 1.53    | 0.28           | 15.9**| 0.86    | 0.79   | 50.0 | 70.0 |
| Pb (mg kg\(^{-1}\))     | 10.2    | 2.50    | 0.51           | 13.7**| 2.10    | 0.69   | 100.0| 180.0|
| Zn (mg kg\(^{-1}\))     | 33.9    | 6.73    | 0.69           | 53.3**| 7.97    | 1.08   | 300.0| 450.0|
| PLI                     | 0.65    | 1.06    |                |       |         |        |      |      |

SD - Standard Deviation; CF - contamination factor; PLI - pollution load index; * (p < 0.05) and ** (p < 0.01) indicate statistical difference (ANOVA; F-test) between mean values of metals in F1 and F2.

Contents of heavy metals in the soil of the study area were higher at F2 farm, downstream of the city of Montes Claros, when compared with those reported by Gebrekidan et al. (2013) for Cr (31.02 mg kg\(^{-1}\)), Pb (3.27 mg kg\(^{-1}\)), and Zn (51.83 mg kg\(^{-1}\)), except Cu (25.25 mg kg\(^{-1}\)) and Ni (26.00 mg kg\(^{-1}\)) in areas irrigated with water of the Ginfel River, polluted by wastewater produced in Sheba Tannery, Ethiopia. The Cr content reported by Chary et al. (2008) (33.00 ± 12.00 mg kg\(^{-1}\)) and by Singh et al. (2010) (19.21 ± 3.26 mg kg\(^{-1}\)) was also lower than the mean value found in F2. However, no heavy metal detected in both farms (F1 and F2) featured values
higher than international and Brazilian limits recommended for soils (CONAMA, 2009; Ewers, 1991) (Table 2).

CF values indicated that Cu, Ni, and Pb in both farms (F1 and F2) and Cr and Zn in F1 farm presented low level of contamination. Cr and Zn in F2 and As in both farms (F1 and F2) showed a moderate level of contamination, according to CF (Hakanson, 1980). PLI indicates the existence of heavy-metal pollution in soil irrigated with the water of Vieira River downstream of Montes Claros, despite the its value in F2 being very close to the limit (PLI = 1) that indicates whether or not there is pollution (Varol, 2011) (Table 2).

Mean values of geaccumulation index ($I_{geo}$) in the soils of both farms (F1 and F2) accounted for the following descending order: As > Cr > Zn > Cu > Pb > Ni. The As in F2 was the only heavy metal that featured positive $I_{geo}$ values, indicating the presence of pollution. $I_{geo}$ value for As in F2 was 0.74, which characterizes the soil in that area as unpolluted to moderately polluted (Table 3). $I_{geo}$ values for Cr, Cu, Ni, Pb, and Zn in both farms (F1 and F2) and for As in F2 were negative, indicating absence of pollution in soils (Table 3).

Table 3. Assessment of geoaccumulation indices ($I_{geo}$) and potential ecological risk for pollution by heavy metals in samples of soil irrigated with Vieira River water in two farms (F1 and F2) (n = 8).

| Sites       | $I_{geo}$ | Mean |
|-------------|-----------|------|
|             | As        | Cr   | Cu  | Ni  | Pb   | Zn   |
| F1 Farm     | -0.10     | -0.78| -1.24| -2.44| -1.56| -1.13| -1.21|
| F2 Farm     | 0.74      | -0.31| -0.90| -0.92| -1.13| -0.48| -0.50|
| Mean        | 0.32      | -0.54| -1.07| -1.68| -1.34| -0.80|

| Sites       | $E_r$ | RI     | Risk level |
|-------------|-------|--------|------------|
|             | As    | Cr    | Cu   | Ni   | Pb   | Zn   |          |          |
| F1 Farm     | 13.98 | 2.21  | 3.85 | 1.25 | 1.31 | 6.97 | 29.56    | Low      |
| F2 Farm     | 25.04 | 3.08  | 4.85 | 3.57 | 1.77 | 10.96| 49.26    | Low      |
| Mean        | 19.51 | 2.64  | 4.35 | 2.41 | 1.54 | 8.96 | 39.41    | Low      |

RI – potential ecological risk index; and $E_r$ – ecological risk factor of certain heavy metal.

Mean $E_r$ of each analyzed heavy metal accounted for the following descending order: As > Zn > Cu > Cr > Ni > Pb. All analyzed metals (As, Cr, Cu, Ni, Pb, and Zn) featured low ecological risk in both farms (F1 and F2), according to the $E_r$ classification proposed by Hakanson (1980). Nevertheless, $E_r$ values were higher in F2, which indicates increased ecological risk in soils irrigated with water of Vieira River downstream of Montes Claros (Table 3). RI values were much lower than 110 in all sampling sites, indicating a low potential ecological risk. The As significantly contributed to the environmental RI, which can be attributed to the effect of anthropogenic activities. The As is used as preservative for leather and wood, additive for metal alloys, and in herbicides (Halli et al., 2014; Thuong et al., 2013). RI value was also higher at F2 (Table 3).

3.3. Heavy metal contamination in plants irrigated with water of Vieira River

Contents of As and Zn in Brachiaria mutica plants in F2 were significantly higher (p < 0.05) than those found in Pennisetum purpureum plants in F1. On the other hand, Cu featured higher concentration (p < 0.01) in Pennisetum purpureum plants (F1). The mean content of heavy metals in plants of both farms (F1 and F2) accounted for the following descending order: Zn > Cu > Ni > As. Cr and Pb were not detected in forage grasses in both farms (F1 and F2) (Table 4).
Table 4. Content of heavy metals (mg kg$^{-1}$) in forage grasses irrigated with Vieira River water in two farms (F1 and F2) and transfer factor (TF) (n = 8).

| Metals | F1 Farm | F2 Farm | Normal intervals$^a$ | Agronomic crop tolerance$^b$ | Widespread values of TF$^c$ |
|--------|---------|---------|----------------------|-----------------------------|-----------------------------|
|        | Mean    | SD     | TF       | Mean    | SD     | TF       |           |                |               |
| As (mg kg$^{-1}$) | 0.22** | 0.04   | 0.70     | 0.02    | 0.48   | 1 - 1.7 | -          | 0.01 - 0.1    |
| Cr (mg kg$^{-1}$)  | 0.00   | 0.00   | 0.00     | 0.00    | 0.00   | 0.1 - 0.5 | 2          | 0.01 - 0.1    |
| Cu (mg kg$^{-1}$)  | 7.63** | 0.68   | 0.51     | 4.60    | 1.18   | 0.23     | 5 - 30     | 5 - 20        | 0.1 - 1    |
| Ni (mg kg$^{-1}$)  | 0.50   | 0.35   | 0.10     | 1.05    | 1.11   | 0.07     | 0.1 - 5    | 1 - 10        | 0.1 - 1    |
| Pb (mg kg$^{-1}$)  | 0.02   | 0.05   | 0.00     | 0.34    | 0.96   | 0.03     | 5 - 10     | 0.5 - 10      | 0.01 - 0.1 |
| Zn (mg kg$^{-1}$)  | 50.77**| 2.74   | 1.54     | 69.92   | 25.11  | 1.32     | 27 - 150   | 50 - 100      | 1 - 10    |

SD – Standard deviation; TF – transfer factor; a – (Kabata-Pendias and Pendias, 2001); b - (Kabata-Pendias and Pendias, 1999; MacNicol and Beckett, 1985); c - (Kloke et al., 1984). * (p < 0.05) and ** (p < 0.01) indicate statistical difference (ANOVA; F-test) between mean values of metals in F1 and F2.

No metal exceeded the normal intervals (Kabata-Pendias and Pendias, 2001) or the agronomic crop tolerance values (Kabata-Pendias and Pendias, 1999; 2001; MacNicol and Beckett, 1985) (Table 4), including Cu and Ni, which featured concentrations in the water higher than the limits recommended by the Brazilian legislation for use of water in irrigation (CONAMA, 2005) (Table 1).

The metal content in forage grasses depends not only on the content of metals in the soil, but also on the efficiency of metal transfer from the soil to the plants. This efficiency can be estimated by TF, which is one of the key indicators of human exposure to metals through the food chain (Cao et al., 2014), and is often used to study environmental pollution (Khan et al., 2013). TF values in both farms (F1 and F2) accounted for the following descending order: Zn > As > Cu > Ni. The highest TF values indicated that Zn and As were easily transferred from the soil to forage grasses when compared with other metals considered in this study. The As was the only metal accounting for TF higher than the widespread values (Kloke et al., 1984) (Table 4).

The As mobility in the soil depends on pH, redox potential, and type and amount of adsorbents (oxides and hydroxides of Fe$^{3+}$, Al$^{3+}$, Mn$^{3+}$, or Mn$^{4+}$, humic substances, and clay minerals) present in the soil (Bissen and Frimmel, 2003; Oscarson et al., 1983). In addition, the amount of As adsorbed in the soil significantly decreased in the presence of organic matter (Bissen and Frimmel, 2003; Redman et al., 2002). Therefore, soil fertilization practices with animal manure at F1 farm, and use for irrigation of Vieira River water contaminated with sewage from Montes Claros at farm F2, contributed to the increase in the organic matter content in soils of these two farms (F1 and F2). This greater amount of organic matter in the soil can result in greater mobility of As (Bissen and Frimmel, 2003; Redman et al., 2002), which corroborates the fact that, in this study, such element featured TF values higher than the widespread values (Table 4).

TF values of heavy metals were lower in F2 when compared with the values in F1 (Table 4), although there is higher content of heavy metal in the F2 soil (Table 2). This can be attributed to frequent flooding and drying of soils in F2 promoted by the flood irrigation system. Xie and Huang (1998) reported that water saturation and drying of soils, promoted by flooded rice irrigation, considerably increased the soil redox potential and, therefore, reduced As solubility and availability.
4. CONCLUSION

Irrigation of agricultural lands with polluted water from Vieira River has increased the concentration of As and Ni in water; of As, Cr, Cu, Ni, Pb, and Zn in the soil; and of As and Zn in forage grasses. Pollution load index (PLI) and contamination factor (CF) indicated the existence of pollution and moderate contamination by As, Cr, and Zn in the soil of the farm downstream of the city of Montes Claros (F2), respectively.

Values of the indicators of ecological risk (Er and RI) observed in the soil of both farms (F1 and F2) showed low ecological risk. However, these indicators were higher in the F2, which indicates increased ecological risk in soils irrigated with water of Vieira River downstream of Montes Claros.

Surface flood irrigation may have contributed to the lower transfer factor (TF) of heavy metals found in the F2 farm; however, the use of this irrigation system in pasture areas is worrisome, because the direct contact of the animal with polluted water from Vieira River may facilitate the contamination of milk or meat of these animals by heavy metals or other inorganic or biological compounds.

Furthermore, there must be greater awareness of contamination of forage grasses and soil by As, since this was the only metal that featured transfer factor (TF) higher than the widespread values found in literature and positive geoaccumulation index (Igeo), indicative of anthropogenic pollution.

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