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Isoprene photo-oxidation products quantify the effect of pollution on hydroxyl radicals over Amazonia

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Nitrogen oxides (NOx) emitted from human activities are believed to regulate the atmospheric oxidation capacity of the troposphere. However, observational evidence is limited for the low-to-median NOx concentrations prevalent outside of polluted regions. Directly measuring oxidation capacity, represented primarily by hydroxyl radicals (OH), is challenging, and the span in NOx concentrations at a single observation site is often not wide. Concentrations of isoprene and its photo-oxidation products were used to infer the equivalent noontime OH concentrations. The fetch at an observation site in central Amazonia experienced varied contributions from background regional air, urban pollution, and biomass burning. The afternoon concentrations of reactive nitrogen oxides (NOx), indicative of NOx exposure during the preceding few hours, spanned from 0.3 to 3.5 parts per billion. Accompanying the increase of NOx concentration, the inferred equivalent noontime OH concentrations increased by at least 250% from 0.6 × 106 to 1.6 × 107 cm−3. The conclusion is that, compared to background conditions of low NOx concentrations over the Amazon forest, pollution increased NOx concentrations and amplified OH concentrations, indicating the susceptibility of the atmospheric oxidation capacity over the forest to anthropogenic influence and reinforcing the important role of NOx in sustaining OH concentrations.

INTRODUCTION

Earth’s atmosphere is an oxidizing medium that drives organic molecules toward carbon dioxide, and oxidation by hydroxyl radicals (OH) initiates most of these reactions (1, 2). Oxidation by OH radicals also leads to the production of many secondary pollutants that affect human health and climate, such as organic particulate matter and ozone. In relation to OH concentrations, the NOx family, defined as including nitric oxide (NO) and nitrogen dioxide (NO2), has two roles (2). The chemistry is illustrated in Fig. 1A. On the one hand, NO reacts with hydroperoxyl radicals (HO2) and organic peroxy radicals (RO2) to produce OH catalytically in the presence of sunlight, thereby enhancing OH concentrations. On the other hand, at higher NOx concentrations, the direct reaction of NOx with NO serves as an OH sink. Hydroxyl radical concentrations thus increase as NOx concentrations increase below a threshold NOx concentration, and they decrease as NOx concentrations increase above this threshold. The net result is that OH concentrations follow a bell curve with respect to NOx concentrations, as illustrated in Fig. 1B (2, 3).

The applicability of this classical understanding of OH–NOx chemistry, as represented by the bell curve, to atmospheric conditions is under challenge based on field measurements of OH concentrations in regions where volatile organic compounds (VOCs) are abundant. A meta-study examined the dependence of OH concentrations across a broad range of NOx conditions (4). The observations ranged from tropical forests in South America and Southeast Asia (5, 6) to deciduous forest in the United States and rural area in China (7–9) and to polluted metropolitan regions of New York City, Beijing, Tokyo, and Mexico...
City (10–13). Above a threshold NO\textsubscript{x} concentration, OH concentrations decreased with increasing NO\textsubscript{x} concentration, as expected. However, below the threshold NO\textsubscript{x} concentration, reported OH concentrations in many regions were unexpectedly high, and collectively, they appeared to be independent of NO\textsubscript{x} concentration, as illustrated by the horizontal line in Fig. 1B.

Possible mechanisms for maintaining elevated OH concentrations under atmospheric conditions with below-threshold NO\textsubscript{x} concentrations have been considered (7, 14–16). In regions where VOCs are abundant, especially over and downwind of forests, the major fate of OH is reaction with biogenic VOCs to produce RO\textsubscript{2}. Possible reactions of RO\textsubscript{2} radicals to regenerate OH and thereby maintain OH concentrations have been suggested, especially for forested regions dominated by isoprene emissions, such as OH production from the reaction of HO\textsubscript{2} with isoprene-derived RO\textsubscript{2} radicals (ISOPOO) (14, 17), as well as OH release following ISOPOO isomerization (16, 18). Even so, these mechanisms appear insufficient to quantitatively explain the reported high OH concentrations at low NO\textsubscript{x} concentrations (14, 17, 18). Alternatively, the accuracy of reports of high OH concentrations is uncertain because of the possibility of interferences in the underlying measurement technique of laser-induced fluorescence (LIF) (9, 19, 20). For instance, Mao et al. (9) attributed 40 to 60% of the nominal OH signal observed in a California forest to interferences. After correction, the data fell on the bell curve, instead of on the horizontal line of Fig. 1B (4). To date, there is no consensus on whether similar interferences may have occurred in the other measurements included in the meta-study. The LIF instruments are custom-built in variable configurations by different research groups. Another uncertainty related to the meta-study is that the span in NO\textsubscript{x} concentration of each underlying study was usually not wide, so no individual study closely tested the response of OH concentration to NO\textsubscript{x} concentration across a broad range. Instead, the analysis in the meta-study was based on normalization of observations across multiple sites, different instruments, and differing atmospheric conditions. In short, possible new mechanisms for OH production and recycling under atmospheric conditions and possible artifacts in the underlying data sets all remain to be reconciled.

Herein, a complementary approach based on isoprene photo-oxygenation products is presented for mapping the dependence of OH concentration on NO\textsubscript{x} concentration for atmospheric conditions over an isoprene-dominated forested environment. Precedent approaches for estimating OH concentrations include the use of a range of OH-reacting trace species, as well as OH reaction products and their ratios (3, 21–24). Isoprene is the dominant VOC emitted to the atmosphere from many forests (25), and reaction with OH is its primary loss pathway (26). As OH concentration increases, the concentration C\textsubscript{ISOP} of isoprene decreases, and the sum concentration C\textsubscript{PROD} of its oxidation products increases, provided that other factors, such as reaction time, ozone concentration, and isoprene emission rates, are unchanged. The presentation herein develops an analysis to use the afternoon parent-to-product concentration ratio C\textsubscript{ISOP}/C\textsubscript{PROD} to infer equivalent noon-time OH concentration within an air mass during the preceding daylight hours. The analysis must account for several factors, in addition to OH concentration, that influence C\textsubscript{ISOP}/C\textsubscript{PROD}.

Data sets were recorded at a single site in the Observations and Modeling of the Green Ocean Amazon (GoAmazon2014/5) Experiment in central Amazonia during the wet and dry seasons of 2014 (27). At different times, the sampling site received mixtures to various degrees of unpolluted regional background air, air influenced by regional fires (especially in the dry season), or air that had passed over the nearby city of Manaus, an urban region of 2 million inhabitants and a strong regional source of NO\textsubscript{x}. The site was 4 to 6 hours downwind from Manaus for the typical prevailing easterlies associated with trade winds (27, 28). As a result, the GoAmazon2014/5 data sets spanned a great breadth of upwind NO\textsubscript{x} chemistry at a single observation site (29). The range in measured isoprene concentrations was similar to that of the forested sites of the meta-study (4). The GoAmazon2014/5 data sets were used in the analysis herein to strongly test for a bell curve versus a horizontal line in the response of OH concentration to NO\textsubscript{x} concentration in an isoprene-dominated forested environment (Fig. 1B).

**RESULTS**

The median and interquartile ranges of measured C\textsubscript{ISOP} and C\textsubscript{PROD} are plotted by hour in Fig. 2 (A and B). The dry and wet season data sets are represented by red and blue colors, respectively. The product concentration C\textsubscript{PROD} represents the sum concentration of methyl vinyl ketone (MVK; C\textsubscript{5}H\textsubscript{4}O), methacrolein (MACR; C\textsubscript{5}H\textsubscript{6}O), and isoprene-derived hydroperoxide isomers [including (1,2)-ISOPOOH and (4, 3)-ISOPOOH, hereafter ISOPOOH; C\textsubscript{6}H\textsubscript{10}O\textsubscript{3}]. MVK and MACR are the major OH-reaction products of isoprene along NO-mediated pathways (17). The

![Fig. 2. Hourly variation of VOC concentrations. (A) Isoprene concentration C\textsubscript{ISOP} and (B) sum concentration C\textsubscript{PROD} of isoprene oxidation products. The vertical dashed gray lines demarcate local sunrise, noon, and sunset (UTC less 4 hours). Data are shown in the wet and dry seasons in blue and red colors, respectively. The solid line and shaded regions, respectively, represent the median and interquartile ranges of the data sets for each hour of the day. The two black dashed lines in (A) show the simulated increase of C\textsubscript{ISOP} from sunrise to midafternoon using Eq. 2 for C\textsubscript{ISOP,0} values of 0.35 and 0.6 ppb.](http://advances.sciencemag.org/Downloaded from)
ISOP species are the major products along HO₂-mediated pathways (30). Although the relative importance of the NO and HO₂ pathways varies with NO₂ concentration (29), the summed production yield of MVK, MACR, and ISOPOH is approximately 70% across the atmospherically relevant range of NO₂ concentrations (30).

Figure 2 (A and B) shows that before sunrise, ISOP and PROD were <20 and <10% of their respective afternoon values. At sunrise, the forest began to emit isoprene, and photochemical production of OH also began. Isoprene had a lifetime of 2.8 hours against OH attack for \( \text{C}_\text{OH} = 1.0 \times 10^9 \text{ cm}^{-3} \), whereas the lifetimes of products against chemical loss were up to five times higher. In line with these processes, Fig. 2 (A and B) shows that at sunrise, ISOP increased, followed shortly thereafter by increases in PROD. Isoprene also reacted with photochemically produced O₃, but this pathway was typically one order of magnitude less important than the OH pathway for the prevailing atmospheric conditions (26). The resulting MVK and MACR was a minor contribution to the overall production of these species. In addition to chemical loss and production, ISOP and PROD were also affected by dry and wet deposition, as well as by entrainment of air from above the boundary layer. The analysis herein focuses on afternoon concentrations ISOP and PROD because OH oxidation has the greatest impact on ISOP and PROD at this time of the day. In addition, the possibility of intercepting a polluted air mass from Manaus was highest in the afternoon, given that the research site was 4 to 6 hours downwind of this pollution source for typical winds following sunrise (28).

Scatterplots for the afternoon hours (13:00 to 16:00 local time) of ISOP, PROD, and PROD/ISOP in relation to the sum concentration \( \text{C}_\text{NO}_x \) of reactive nitrogen species (NOₓ) are shown in Fig. 3, A to C, respectively. In addition to the NOₓ species, the NOₓ family also included reservoir forms of oxidized nitrogen, such as HNO₃, which were products of atmospheric reactions of NOₓ (3). A few hours downwind of the Manaus source region at the observation site, \( \text{C}_\text{NO}_x \) remained a semiconserved quantity, whereas \( \text{C}_\text{NO}_x \) had significantly decreased. Atmospheric conversion of NOₓ to NO₂ also occurred for air masses downwind of regional biomass burning. Therefore, for the analysis, \( \text{C}_\text{NO}_x \) was used to represent the integrated intensity of NOₓ chemistry before arrival of an air mass at the observation site (29, 31). The ratio \( \text{C}_\text{NO}_x/\text{C}_\text{NO}_x \) had an interquartile variability of 0.30 to 0.44 across the study period, supporting the use of the simplified representation \( \text{C}_\text{NO}_x \).

As shown in Fig. 3, the hourly averaged afternoon values of \( \text{C}_\text{NO}_x \) spanned more than one order of magnitude from 0.3 to 3.5 parts per billion (ppb). The interquartile range of \( \text{C}_\text{NO}_x \) in the afternoon hours was 0.7 to 1.2 ppb in the wet season and 1.3 to 2.0 ppb in the dry season. As a reference point, \( \text{C}_\text{NO}_x \) for regional background conditions was 0.46 ± 0.26 ppb (32). Concentrations above 1 ppb at the observation site indicated the effects of Manaus pollution or regional biomass burning (29, 31).

The scatterplots of Fig. 3 illustrate the following findings. ISOP did not correlate with \( \text{C}_\text{NO}_x \) (Spearman’s rank \( r = 0.4 \) and correlation coefficient \( r = -0.07 \); Fig. 3A), PROD did correlate with \( \text{C}_\text{NO}_x \) (\( P < 10^{-4} \); Fig. 3B), yet the data were scattered (\( r = 0.4 \)). By comparison, the ratio \( \text{C}_\text{PROD}/\text{C}_\text{ISOP} \) correlated tightly with \( \text{C}_\text{NO}_x \) (\( P < 10^{-4} \) and \( r = 0.6 \); Fig. 3C). The median of \( \text{C}_\text{PROD}/\text{C}_\text{ISOP} \) increased from 0.4 to 1.0 as \( \text{C}_\text{NO}_x \) changed from below 0.5 to above 2 ppb. A high value of \( \text{C}_\text{PROD}/\text{C}_\text{ISOP} \) was never observed for a low value of \( \text{C}_\text{NO}_x \) (Fig. 3C).

An unclear trend in the concentration plots on the one hand (Fig. 3, A and B) compared to a clear trend in the ratio plot on the other hand (Fig. 3C) with respect to \( \text{C}_\text{NO}_x \), can be explained by the large temporal variability in isoprene emissions (22). This variability was independent of \( \text{C}_\text{NO}_x \) and, thus, confounded direct relationships between \( \text{C}_\text{ISOP} \) and \( \text{C}_\text{NO}_x \), \( \text{C}_\text{PROD} \) and \( \text{C}_\text{NO}_x \). By comparison, the ratio \( \text{C}_\text{PROD}/\text{C}_\text{ISOP} \) largely compensated the variability in isoprene emissions and thereby revealed differences in atmospheric oxidation. For these reasons, \( \text{C}_\text{PROD}/\text{C}_\text{ISOP} \) was used in the further analysis herein, focusing on understanding and quantifying the effects of pollution on the atmospheric oxidation cycle over central Amazonia.

**DISCUSSION**

**Relating OH concentration to the ratio \( \text{C}_\text{PROD}/\text{C}_\text{ISOP} \)**

A model of isoprene photochemistry over the course of 1 day was constrained by the observed values of \( \text{C}_\text{PROD}/\text{C}_\text{ISOP} \) to estimate the equivalent noontime OH concentration for the preceding daytime.
which is a commonly used simplification of Eq. 1 to relate an observed

\[ \frac{dC_i}{dt} = y_i(NO) k_{ISOP,OH} C_{OH}(t) C_{ISOP}(t) + \]

\[ y_{1,OH} k_{ISOP,OH} C_{OH}(t) C_{ISOP}(t) - k_t C_i(t) \]

(1)

where \( y_i \) is the production yield of product species \( i \) from the reaction between hydroxyl radical and isoprene. It varies according to the fate of ISOPOO, which is largely controlled by the NO concentration and hence susceptible to pollution. \( C_{OH}(t), C_{ISOP}(t), \) and \( C_{O_3}(t) \), where \( t \) represents time, are the concentrations of hydroxyl radical, isoprene, and ozone, respectively, all of which vary strongly with time of day. \( k_{ISOP,OH} \) and \( k_{ISOP,OH} \) are the reaction rate constants of isoprene with hydroxyl radical and ozone, respectively. \( y_{1,OH} \) is the production yield of product species \( i \) from the reaction between ozone and isoprene (30). \( k_t \) is a composite, pseudo-first-order loss coefficient of species \( i \), given by \( k_t(t) = k_{1,OH} C_{OH}(t) + k_{1,O_3} C_{O_3}(t) + k_{1,e} + k_{de} \) for bimolecular reaction between species \( i \) and \( OH \) or \( O_3 \), entrainment mixing with air above the boundary layer (en), and deposition to the planetary surface (de). Table S1 lists values of \( y_{1,OH}, k_{1,OH}, \) and \( k_{1,OH} \), as well as typical values of \( k_{1,e} \) and \( k_{de} \) for central Amazonia, partly reproduced from Liu et al. (29). For comparison, a more detailed model for entrainment to couple chemistry and boundary layer dynamics was also explored, and similar results were obtained (see the Supplementary Materials).

The text below first discusses the sequential reaction model (SRM), which is a commonly used simplification of Eq. 1 to relate an observed ratio \( C_{PROD/C_{ISOP}} \) to an inferred OH concentration (21). The conclusion reached is that the SRM is not appropriate for analysis of the collected data set. The text then introduces a complementary approach based on (i) time-dependent approximations of \( C_{PROD}, C_{OH}, \) and \( C_{O_3} \), and (ii) substitution of the production yield \( y_i(NO) \) by an effective production yield denoted by \( y^*_i(C_{ISOP}) \).

Analysis by SRM assumes that isoprene emissions can be taken as a point source (typically an upwind forest) and that isoprene is subsequently oxidized during transport to an observation site. More specifically, isoprene is not emitted into the air parcel along its path of travel, and reaction with OH is the dominant process governing the time evolution of \( C_{ISOP} \) (that is, \( dC_{ISOP}/dt = -k_{ISOP,OH} C_{OH} C_{ISOP} \)). Other assumptions include constant \( C_{OH} \) and \( C_{O_3} \), a fixed \( y_i \), a fixed reaction time \( t \), and negligible entrainment and deposition. For these conditions, Eq. 1 transforms to an explicit relation that allows \( C_{OH} \) to be inferred from measured \( C_{PROD/C_{ISOP}} \) (21). The major underlying SRM assumption of an upwind point source was, however, not applicable to the GoAmazon2014/5 scenario. The observation site was surrounded by forest for hundreds of kilometers, meaning that isoprene was continuously emitted into air parcels throughout transport and that \( C_{ISOP} \) and \( C_{PROD} \) corresponded to the integrated balance between source and loss processes throughout transport.

An alternative scheme to the SRM is developed herein to constrain \( C_{OH} \) based on observed \( C_{PROD/C_{ISOP}} \). The regional area around T3 is approximated as a homogeneous forest representing a perfectly diffuse nonpoint source region of isoprene. Analysis using Google Earth shows that surface forest coverage exceeded 70% in a 100-km radius for the dominating easterlies and northeasterlies in the wind rose and >80% for all directions (33). The radius of 100 km was based on typical wind speeds of 10 to 20 km h⁻¹ and the time period from the start of photochemistry at daybreak to the analysis window in the afternoon. Within this 100-km radius, the forest type and, hence, isoprene emissions varied to some extent (34), and there were also scattered pastures, two large rivers, and the urban area of Manaus. These factors notwithstanding, a homogeneous diffuse source of isoprene emissions was taken as an acceptable approximation for the accuracy of the modeling herein.

Under this approach of a homogeneous source region, a time series of observations at a Eulerian point is fully transformable into a Lagrangian model of a time series of concentrations within an air parcel over the course of a time period (35, 36). Equation 1 describes the transformations of product species \( C_i \) within the Lagrangian parcel, and in the treatment herein, observations constrained the terms. For instance, measurements of the time course of isoprene concentrations directly constrained \( C_{ISOP}(t) \). Steady increases were observed from sunrise to midafternoon (Fig. 2A), and the observed time dependence was represented empirically by the following linear equation

\[ C_{ISOP}(t) = \xi C_{ISOP,0} (1 + t/t_{ISOP}) \]

where \( \xi \) was a daily scaling factor representing the variability in isoprene concentrations for each day, \( C_{ISOP,0} \) was the typical isoprene concentration at time zero (sunrise), and \( t_{ISOP} \) was the typical characteristic time for \( C_{ISOP} \) to double its initial value. A value of 2.5 hours for \( t_{ISOP} \) represented most days. Values of \( C_{ISOP,0} \) of 0.35 and 0.6 ppb were used as approximate values to represent the wet and dry seasons, respectively. Equation 2 is plotted in Fig. 2A as the black dashed lines to represent median \( C_{ISOP}(t) \) from sunrise to midafternoon in the two seasons using \( \xi \) of unity. The analysis below further shows that the quantities \( \xi \) and \( C_{ISOP,0} \) drop out (Eq. 5) so that the actual values are not important.

In regard to \( C_{OH}(t) \), hydroxyl radical concentrations strongly correlate with the photolysis frequency \( J_{OH} \) of ozone for most locations worldwide (37). In the analysis, the daily course of hydroxyl radical concentrations was approximated as an equivalent noontime concentration \( C_{OH,noon} \) modulated by a time-dependent photolysis frequency \( J_{OH} \) of ozone along the \( O^1D \) channel, as follows

\[ C_{OH}(t) = C_{OH,noon} \frac{J_{OH}(t)}{J_{OH,noon}} \]

where \( J_{OH,noon} \) was the peak value occurring at local noon for the equatorial location. The variation of \( J_{OH}(t)/J_{OH,noon} \) followed that of the Master Chemical Mechanism for clear skies at the latitude and longitude of the observation site (fig. S2) (30). The value of \( C_{OH,noon} \) was not known before the analysis, and inferring its value was a main point of the analysis presented herein.

The production yield \( y_i(C_{ISOP}) \) in Eq. 1, written in full form as \( y_i[C_{ISOP} C_{NO}(t)] \), does not fulfill the homogeneity requirement of a Eulerian-Lagrangian transformation because of the decreasing concentration of NO along the path. The observations and analysis presented in the study of Liu et al. (29) for this same observation site showed that the ratio of product concentrations \( C_{ISOP}/C_{MVK+MACR} \) correlated tightly with \( C_{NO} \), and this ratio was further transformed in that study to the effective production ratio \( y^*_{ISOP//OHH^*_{MVK+MACR} \} \) by combining the measurements with kinetic modeling (29). Herein, the production
yield \( y_i[CNO(t)] \) in the Lagrangian framework was substituted by an empirical effective production yield \( y_i'(CNO) \). The relationship \( y_i'(CNO) \) is presented in fig. S3 (see further in the Supplementary Materials). This treatment subsumed the reality of a detailed history of NO exposure within the sampled air parcel along its Eulerian path into an effective behavior, which was empirically quantified by semi-conserved \( C_{NO} \) during the course of Lagrangian time. Although this approach was approximate, the tight relationship of concentration ratio and \( C_{NO} \) observed in the study of Liu et al. (29) supported its use for the analysis herein.

The quantity \( C_{O_3}(t) \) in Eq. 1 was directly constrained by two observations: (i) the increase of ozone concentrations from sunrise to midafternoon (fig. S4A) and (ii) the correlation between \( C_{O_3} \) and \( C_{NO} \) for afternoon time periods (fig. S4B). On the basis of these observations, \( C_{O_3}(t) \) for an air parcel arriving at the observation site in the afternoon was empirically approximated as a function of observed \( C_{NO} \). Specifically, \( C_{O_3}(t) \) was represented by \( C_{O_3}(C_{NO}, t) \), as explained further in the Supplementary Materials. Overall, sensitivity tests presented herein show that the influence of ozone chemistry on \( C_{PROD/CISO} \) and, hence, inferred \( C_{O_{3, noon}} \) was small.

For these treatments of \( C_{ISO}(t), C_{OH}(t), C_{O_3}(t), \) and \( y_i(NO) \), Eq. 1 can be rewritten as follows

\[
\frac{dC_i}{dt} = \begin{cases} 
\left[ y_i'(CNO) \right] k_{ISO,OH} & C_{OH_{noon}} \frac{J_{OH}(t)}{J_{OH,noon}} + \\
y_{i,O_2} k_{ISO,O_3} C_{O_3}(C_{NO}, t) & C_{ISO,0} \left[ 1 + \frac{t}{t_{ISO}} \right] - \\
k_i(t) & C_i(t) \times (C_{ISO,0} - C_{ISO,0})(t)
\end{cases} \quad C_i(0) = 0
\]

The initial concentrations of product species \( i \) are taken as zero [that is, \( C_i(0) = 0 \)] based on the data sets of Fig. 2. The symbol \( \zeta_i \), defined by \( \zeta_i = C_i/C_{ISO} \), is introduced for the product ratio. By definition, \( C_{ISO} = \Sigma_i \zeta_i \). The following equation holds for \( d\zeta_i/dt \)

\[
\frac{d\zeta_i}{dt} = y_i'(CNO) k_{ISO,OH} C_{OH_{noon}} \frac{J_{OH}(t)}{J_{OH,noon}} + \\
y_{i,O_2} k_{ISO,O_3} C_{O_3}(C_{NO}, t) - \left( k_i(t) + \frac{1}{t_{ISO} + t} \right) \zeta_i(t) \quad \zeta_i(0) = 0
\]

Equation 5 is derived in the Supplementary Materials. The scaling factor \( \xi \) and the initial concentration \( C_{ISO,0} \), drop out. Equation 5 captures the behavior shown in Fig. 3 that the ratio \( C_{PROD/CISO} \) is independent of isoprene emissions even as the product concentration \( C_{PROD} \) is not. Equation 5 also suggests that the \( NO_x \) exposure of the air mass, represented by \( C_{NO} \), here, is a key driver of \( C_{ISO} \) under ambient conditions through (i) effects on \( y_i'(CNO) \), (ii) effects on \( C_{O_3}(C_{NO}) \), and (iii) possible effects on \( C_{OH_{noon}}(C_{NO}) \).

The relative sensitivity of the analysis to \( y_i'(CNO) \), \( C_{O_3}(C_{NO}) \), and \( C_{OH_{noon}}(C_{NO}) \) is examined in Fig. 4. The figure shows \( C_{PROD/CISO} \) as a function of \( C_{NO} \) for two fixed values of \( C_{OH_{noon}} \), specifically \( 1 \times 10^6 \) and \( 2 \times 10^6 \) cm\(^{-3}\), using Eq. 5 with and without ozone chemistry. At fixed \( C_{OH_{noon}} \), \( C_{NO} \), affects \( C_{PROD/CISO} \) via the effects on \( y_i'(CNO) \) and \( C_{O_3}(C_{NO}) \) (Eq. 5). Solid lines show results for the full model of Eq. 5, and dashed lines show results for a model that omits isoprene ozonolysis (that is, \( C_{O_3} = 0 \)). For all cases, \( t = 8.5 \) hours, corresponding to 14:30 (local time).

The simulations in Fig. 4 can be compared to the observations in Fig. 3C. They are in agreement that \( C_{PROD/CISO} \) increases for greater \( C_{NO} \). For fixed \( C_{OH_{noon}} \), the relative increase is, however, smaller in the simulation than in the observations. For an increase of \( C_{NO} \) from 0.3 to 3.5 ppb, the simulated increase of \( C_{PROD/CISO} \) is 50 to 60% for fixed \( OH \) concentration, whereas the observed increase is greater than 200%. Changes in \( y_i \) and \( C_{O_3} \) with \( C_{NO} \) thus do not fully explain the observed dependence of \( C_{PROD/CISO} \) on \( C_{NO} \). The implication is that changes in \( C_{OH_{noon}} \) with \( C_{NO} \) are important for explaining the observations. More specifically, \( C_{OH_{noon}}(C_{NO}) \) can be inferred from the data set of \( C_{PROD/CISO}(C_{NO}) \).

To do so, Eq. 5 was used in conjunction with the data set of \( C_{PROD/CISO}(C_{NO}) \) plotted in Fig. 3C to estimate the associated values of equivalent noon-time OH concentration \( C_{OH_{noon}} \). The values are described as equivalent, given the approximation of the diel variation of \( OH(t) \), as well as the path-averaging inherent in \( y_i'(CNO) \). For this analysis, projected values of \( C_{PROD/CISO} \) were calculated by integrating Eq. 5 and summing the resultant concentration ratio \( \zeta_i \) of individual products for processing times of 7.5, 8.5, and 9.5 hours across an array of atmospherically relevant values of \( C_{OH_{noon}} \) and \( C_{NO} \). A lookup matrix of four dimensions (that is, \( C_{PROD/CISO} C_{OH_{noon}} C_{NO} \), and \( t \) was thereby created. A value of \( C_{OH_{noon}} \) was obtained from the matrix based on \( C_{PROD/CISO} C_{NO} \), \( t \) for each data point of Fig. 3C. The values of \( C_{OH_{noon}} \) ranged from \( 0.3 \times 10^6 \) to \( 4 \times 10^6 \) cm\(^{-3}\) across the data set of Fig. 3C. The quartiles were \( 0.8 \times 10^6 \), \( 1.1 \times 10^6 \), and \( 1.4 \times 10^6 \) cm\(^{-3}\).
Entrainment of air aloft accompanying the growth of the convective boundary layer is another important process that can affect the inferred \(C_{OH,noon}\) values. Sensitivity tests were performed to evaluate the treatment embedded in \(k(t)\) of Eq. 1 as compared to a more detailed mixing model. The results, presented in the Supplementary Materials, suggest that \(C_{OH,noon}\) can be underestimated on the order of 20% for higher values of \(C_{NO_x}\), again indicating that the line drawn in Fig. 5 can be an underestimate.

Random measurement errors of \(C_{PROD}, C_{ISOP}\), and \(C_{NO_x}\), and systematic errors in the parameterizations of \(C_{OH}(t), C_{ISOP}(t),\) and \(y_x^*(C_{NO_x})\) were also considered. The methods are presented in the Supplementary Materials (see table S2). As a result, inferred \(C_{OH,noon}\) values are estimated to have a standard error of ±30%. This uncertainty is small relative to the trend of an increase by 250% of \(C_{OH,noon}\) from small to high \(C_{NO_x}\).

In the context of the controversy involving direct OH observations (9, 19, 20), the inference of OH concentrations from measurements of isoprene and its oxidation products, as presented herein, represents an important, albeit indirect, complementary analysis. The result presented in Fig. 5 serves as counterevidence to the conclusion of the earlier meta-study, suggesting that OH concentrations are independent of \(NO_x\) in low-\(NO_x\), high-isoprene environments (4). The OH concentrations obtained herein were 80% lower than those reported using the LIF technique over rainforest in coastal eastern South America for similar isoprene concentrations. By comparison, the OH concentrations inferred by the analysis herein were comparable to concurrent GoAmazon2014/5 observations using chemical ionization mass spectrometry deployed at the same observation site (see the Supplementary Materials). In addition, the OH concentrations fell into the range of OH concentrations inferred from previous airborne measurements over tropical forests in South America (see table S3 and references therein).

In summary, the current study shows that OH concentrations inferred from measurements of isoprene and its oxidation products increased with increasing \(NO_x\) concentrations over a tropical forest in the central Amazon basin, in support of the bell curve response represented in Fig. 1. This result suggests that the oxidation capacity over tropical forests is susceptible to anthropogenic \(NO_x\) emissions. Looking into the future in Amazonia, trends of ongoing deforestation for agricultural purposes and expanding urbanization can be expected to increase demand for electricity and transport (39). Increased \(NO_x\) emissions should be expected on the basis of current technologies. Hydroxyl radical concentrations can be expected to likewise increase, keeping other factors equal, in forested regions affected by anthropogenic pollution in Amazonia. Increased OH concentrations imply changed spatial and temporal oxidation patterns for VOCs emitted from the forest, which can have follow-on effects on visibility, cloud formation, and rainfall based on shifted mass concentration, size distribution, and chemical composition of organic particulate matter (28, 40).

**Fig. 5. Dependence of inferred equivalent noontime OH concentration \(C_{OH,noon}\) on \(NO_y\) concentration.** The gray dots represent \(C_{OH,noon}\) inferred for the individual data points of Fig. 3C. The red crosses represent medians and quartiles after grouping the data points into five equally spaced bins based on logarithmic \(NO_y\) concentrations. The orange line connects the medians of the binned data.

The values for \(C_{OH,noon}\) estimated by this analysis represent effective noontime OH concentrations that the air parcel experienced, embodying the history of oxidative capacity along its path of travel before arrival at the observation site.

**Dependence of \(C_{OH,noon}\) on \(C_{NO_x}\)**

The retrieved \(C_{OH,noon}\) values are plotted versus \(C_{NO_x}\), in Fig. 5. The data points distribute below the 45° line of the plot, indicating that high OH concentrations occurred only for high \(C_{NO_x}\). For further statistics, the \(C_{OH,noon}\) values were grouped by equally spaced logarithmic bins of \(C_{NO_x}\), and quartile and median values were taken for each subset. The median values of \(C_{OH,noon}\) increased monotonically with median values of \(C_{NO_x}\). For the subset of lowest \(C_{NO_x}\), corresponding to a median value of 0.4 ppb, the median value of \(C_{OH,noon}\) was \(0.6 \times 10^6\) cm\(^{-3}\). For the subset of highest \(C_{NO_x}\), corresponding to a median value of 2.6 ppb, the median value of \(C_{OH,noon}\) was \(1.5 \times 10^6\) cm\(^{-3}\). The median value of \(C_{OH,noon}\) thus increased by 250% for a shift in \(C_{NO_x}\) from 0.4 to 2.6 ppb.

Even as there is a clear increasing trend of \(C_{OH,noon}\) for increasing \(C_{NO_x}\) (Fig. 5), there is also a scatter in the data for high \(C_{NO_x}\). The scatter at high \(C_{NO_x}\) could be related to some underlying approximations in the retrieval of \(C_{OH,noon}\) such as the use of \(C_{NO_x}\) to represent integrated \(NO_x\) exposure. The omission from the model of possible wet deposition upward of the observation site might also be a major factor. Although the data were screened for fair weather locally, patchy afternoon precipitation, caused by locally induced thermal convection, frequently occurred in the study region (38). In the case that an air parcel encountered a rainfall event before arriving at the observation site, the more soluble oxidation products of isoprene might be removed preferentially relative to isoprene, and the retrieved \(C_{OH,noon}\) values for these cases would be biased low. The suggestion then is that the line drawn in Fig. 5 is a lower estimate of the increase in \(C_{OH,noon}\) for increasing \(C_{NO_x}\), because of omission of wet deposition in the analysis.
the fetch of air at T3 varied from background conditions of the tropical forest (that is, low NO\textsubscript{3} to polluted conditions under the influence of Manaus and regional biomass burning (that is, higher NO\textsubscript{3} compounds. The product analysis herein was based on the sum concentration of NO\textsubscript{3}, CO\textsubscript{3}.fig. S4. Observation and simulation of ozone concentration CO\textsubscript{3}.Liu et al section S8. Comparison of section S6. Sensitivity tests regarding entrainment process section S4. Derivation of Eq. 5 in the main text figures. 1. H. Levy II, Normal atmosphere: Large radical and formaldehyde concentrations predicted. 2. D. H. Ehhalt, Photooxidation of trace gases in the troposphere Plenary Lecture. 36–516 (2014). 3. J. Peeters, T. L. Nguyen, L. Vereecken, HO\textsubscript{2} chemistry in the urban atmosphere of New York City. Atmos. Environ. 71, 4639–4650 (2013). 4. J. D. Crounse, F. Paulot, H. G. Kjaergaard, P. O. 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