Net ecosystem dissolution and respiration dominate metabolic rates at two western Atlantic reef sites

Melissa Meléndez,1* Joseph Salisbury,2 Dwight Gledhill,3 Chris Langdon,4 Julio M. Morell,5,6 Derek Manzello,7 Adrienne Sutton8

1Department of Oceanography, University of Hawai‘i at Mānoa, Honolulu, Hawai‘i
2Ocean Processes Analysis Laboratory, Department of Earth Sciences, University of New Hampshire, Durham, New Hampshire
3Ocean Acidification Program, NOAA, Silver Spring, Maryland
4Rosenstiel School of Marine and Atmospheric Science, University of Miami, Miami, Florida
5Department of Marine Sciences, University of Puerto Rico, Mayagüez, Puerto Rico
6Caribbean Coastal Ocean Observing System, Magueyes Island, Lajas, Puerto Rico
7Satellite Oceanography and Climate Division, Center for Satellite Applications and Research, NOAA, College Park, Maryland
8Pacific Marine Environmental Laboratory, NOAA, Seattle, Washington

Abstract

Ocean acidification is changing surface water chemistry, but natural variability due to nearshore processes can mask its effects on ecosystem responses. We present an approach of quantitatively resolving net ecosystem metabolism from an array of long-term time series stations, offering perhaps the longest record of such processes over a reef to date. We used 8 and 6 yr of in situ, high-quality frequency observations to characterize the changes in dissolved inorganic carbon and oxygen in La Parguera, Puerto Rico and Cheeca Rocks, Florida, respectively. Net respiration and net dissolution are the dominant metabolic processes at both systems, with a narrow window of ~4 months under net calcification. The annual mean net ecosystem calcification (NEC) rates for La Parguera (~0.68 ± 0.91 kg CaCO3 m⁻² yr⁻¹) and Cheeca Rocks (~0.48 ± 0.89 kg CaCO3 m⁻² yr⁻¹) were on the lower end of typical NEC ranges determined for other reef areas using chemistry- and census-based approaches. At Cheeca Rocks, 53% of the variance in NEC can be explained by net ecosystem production (NEP) and 30% by aragonite saturation state (Ωarag). At La Parguera, NEC is primarily driven by changes in NEP. The linear relationship between NEC and NEP showed a significant slope (± standard error) of 1.00 ± 0.005 and 0.88 ± 0.04 for La Parguera and Cheeca Rocks, respectively. These results suggest that NEP appears to play a prominent role on NEC, and Ωarag probably is not the most informative measure to monitor when attempting to resolve the long-term impacts of ocean acidification.

Surface ocean carbon chemistry is currently changing owing to the burning of fossil fuels (Friedlingstein et al. 2019). The increase in atmospheric carbon dioxide (CO₂atm) from anthropogenic activities has resulted in an increase in CO₂ in surface seawater (CO₂sw) and ocean acidification (Bates et al. 2012). Physical, chemical, and biological processes that influence the surface ocean concentration of CO₂ in nearshore areas complicate our understanding of the carbonate system (Duarte et al. 2013). Models for open-ocean waters in tropical systems (Gledhill et al. 2008) have helped characterize ocean acidification trends in open-ocean waters and have informed the debate about future ocean acidification conditions in coral reef areas (Cyronak et al. 2014). Nevertheless, current and future trends in ocean acidification are still unknown because of the difficulty associated with separating nearshore biogeochemical processes from total ocean changes in surface pH.

Direct observations at coral reef sites in the western Atlantic show significant changes in the coastal surface seawater carbonate chemistry relative to the global open ocean because of enhanced biological and temperature dynamics near the coast (Bates et al. 2012). It is critical to examine areas where substantial changes in carbonate chemistry are expected to occur (Andersson et al. 2019), especially those that are readily accessible or have ongoing related monitoring programs (Sutton et al. 2014, 2019). Coral reef areas in the western Atlantic, where surface waters are likely to be affected by ocean acidification and to warm sooner and more intensely than other reef areas,show significant changes in the coastal surface seawater carbonate chemistry relative to the global open ocean because of enhanced biological and temperature dynamics near the coast (Bates et al. 2012). It is critical to examine areas where substantial changes in carbonate chemistry are expected to occur (Andersson et al. 2019), especially those that are readily accessible or have ongoing related monitoring programs (Sutton et al. 2014, 2019). Coral reef areas in the western Atlantic, where surface waters are likely to be affected by ocean acidification and to warm sooner and more intensely than other reef areas,
regions, have experienced significant reductions in live coral cover, causing a shift from reefs dominated by scleractinian corals to algae-dominated reefs (Gardner et al. 2003). Characterization of the ecosystem metabolic response is essential if we are to understand the sensitivity of nearshore coral reefs to ocean acidification.

Biological processes driving CO2 changes on diurnal and seasonal timescales include net ecosystem production (NEP) and net ecosystem calcification (NEC) from all autotrophic, heterotrophic, biotic, and abiotic components of the system. NEP is the difference between gross primary production and ecosystem respiration, and NEC is the difference between calcification and dissolution, as described in the following simplified equations (Gattuso et al. 1998):

\[
\text{CO}_2 + \text{H}_2\text{O} \xrightarrow{\text{photosynthesis}} \text{CH}_2\text{O} + \text{O}_2, \quad \text{respiration} \\
\text{Ca}^{2+} + \text{CO}_3^{2-} \xrightarrow{\text{calcification}} \text{CaCO}_3, \quad \text{dissolution}
\]

Calcification and organic respiration lead to increases in seawater partial pressure of CO2 ($p\text{CO}_2\text{sw}$) and dissolved inorganic carbon (DIC) that can potentially reduce both the ocean pH and the ocean’s ability to remove CO2 from the atmosphere. The uptake of DIC via photosynthesis and the increase in total alkalinity (TA) via calcium carbonate (CaCO3) dissolution both lower the surface $p\text{CO}_2\text{sw}$ concentration and increase the buffering capacity of seawater.

Several methods are commonly used to measure NEP and NEC processes in nearshore surface waters. For example, net calcification and production measurements typically adopt various Lagrangian or Eulerian techniques requiring high-temporal resolution of discrete TA and DIC measurements in seawater (Albright et al. 2015). Recently, in situ autonomous methods have been proposed that offer great promise but currently remain cost-prohibitive and have a limited operating duration of a few weeks, at most, due to biofouling and sensor drift (Yates and Halley, 2003; Takeshita et al. 2016). Modeling NEP and NEC processes using carbon balance approaches have proven successful in surface open-ocean settings (Gruber et al. 1998; Shadwick et al. 2011). The adoption of this technique has been limited in coastal zones partially because of the limited number of direct high-frequency observations available to quantitatively separate processes affecting CO2 and capture the fast timescales of coastal processes. We identified only two coral reef time series in the western Atlantic (Sutton et al. 2019) with temporal spans that were adequate for detecting seasonal trends and were accompanied by auxiliary data to quantitatively estimate the magnitude of NEP and NEC based on the temporal evolution of oxygen ($O_2$) and DIC.

In this study, an existing mass balance approach for CO2 (Gruber et al. 1998) is implemented, refined, and expanded to decouple organic and inorganic metabolism (NEP and NEC) using $p\text{CO}_2$, DIC, and $O_2$ as proxies. We apply this diagnostic technique to a one-dimensional (1D) mass balance model using CO2 climate quality (Newton et al. 2015) and high-frequency (3-h) carbonate observations spanning 6–8 yr from two shallow nearshore reef areas in Florida and Puerto Rico. The model assumes an open system and well-mixed water column to provide an integrated assessment of biogeochemical variability over time within a water parcel close to the buoy (<5 km). Our ecosystem metabolic rates integrate over extended periods (months to years) and are representative of multiple habitats. We compare and contrast our results with previous studies in Florida and Puerto Rico using chemistry- and census-based approaches. We also evaluated the correlation between NEC and NEP to provide information about the role of organic metabolism in NEC and our ability to define geochemical thresholds at which dissolution exceeds calcification (Andersson and Gledhill 2013)

Given the decline in carbonate production reported for the Caribbean reefs (Perry et al. 2013), we expect that both locations exhibit low NEC that is closely linked to low coral cover, sediment dissolution, high temperatures and organic carbon inputs.

**Methods**

**Study sites**

The buoy in Puerto Rico is located in La Parguera Natural Marine Reserve at the southwest end of the Enrique mid-shelf reef. The buoy in Florida is in the upper Florida Keys National Marine Sanctuary over Cheeca Rocks, an isolated nearshore patch reef. Both buoys are in shallow areas (3 m depth) approximately 2 km off the coast (Fig. 1).

The spatiotemporal scales of the La Parguera buoy cover approximately 0.5 km of the Enrique forereef over a semidiurnal cycle (12 h) due to the low current speed observed in the area (0.05 km h$^{-1}$) and the predominant southwest direction (Fig. 1). Forereef benthic calcifiers (Supporting Information Table S1) include octocorals (26%), hard corals (10–11%), and crustose coralline algae (1%). No secondary (calcareous encrusters) or sediment (calcareous algae such as *Halimeda* or benthic foraminifera) producers were found during multiple visual surveys (Moyer et al. 2012). The percentage of the total forereef area covered by benthic primary producers (e.g., cyanobacteria, macroalgae, and turf algae) is 17%. Carbonate sand, rubble, and rock (35%) are the dominant abiotic components in this area. *Rhizophora mangle* and *Laguncularia racemosa* mangroves have colonized the Enrique reef crest. The inorganic nutrient concentrations across the Caribbean and La Parguera coral reefs are low (Meléndez et al. 2020). Riverine influence is at its minimum during the winter and spring, while the Amazon and Orinoco plumes dominate the southeastern Caribbean during the summer and fall (Corredor and
Morell, 2001). The average seawater temperature at La Parguera ranged from 27.1 ± 0.2°C in winter to 29.7 ± 0.3°C in summer.

The predominant current direction at the Cheeca Rocks buoy is southward (Meléndez et al. 2020), with a maximum current speed of 0.3 km h⁻¹ (Fig. 1). The spatiotemporal coverage over a semidiurnal cycle is approximately 3 km. The benthic communities located north of the buoy include hard bottom areas with fine-grained CaCO₃ sediments, some isolated hard and soft corals, and seagrass. Surveys at the Cheeca Rocks buoy site show that macroalgae (46.6%) and hard coral (25.5–33%) dominate the benthic composition, followed by 6.5% sand, 6.1% secondary carbonate producers, and 1–3% soft corals (Supporting Information Table S1). The tidal exchanges can also send pulses of Florida Bay water into inner reef areas through different tidal channels, particularly during the winter (Smith and Lee, 2003). The bay consists primarily of shallow (< 8 m) seagrass and sand environments, with occasional, isolated patch reefs intermixed (Jones, 1977). The average seawater temperature at Cheeca Rocks ranged from 23.4 ± 0.9°C in winter to 30.6 ± 0.5°C in summer.

**Autonomous observations**

The Pacific Marine Environmental Laboratory (PMEL) time series includes a moored autonomous pCO₂ system (MapCO₂) that provides 3-h measurements of both pCO₂air and pCO₂sw, sea surface salinity (SSS) and sea surface temperature (SST), pH, and O₂. The buoys have a seawater–gas equilibrator, a reference gas standard, an infrared gas analyzer, a sea-bird conductivity and temperature recorder (SBE 16 SeaCAT), a SAMI-pH system, and an O₂ optode (Aanderaa 4775) at approximately 1 m depth. The overall uncertainty in autonomous pCO₂sw is <2 μatm (Sutton et al. 2014). We used autonomous measurements from January 2009 (La Parguera) and November 2011 (Cheeca Rocks) to September 2017.

An internal Maxtec™ oxygen sensor (MAX-250+) is inside the CO₂ electronics tube to measure the percentage of oxygen in the air (% O₂) and is used primarily for the MapCO₂ system diagnostic, but has also been applied to studying ocean biological variability (Massaro et al. 2012; Xue et al. 2016). In this study, the autonomous measurements made by the MAX-250+ oxygen were post-calibrated using the slope and offset of the
linear correlation with the higher-quality O₂ optode measurements (Meléndez et al. 2020) to derive daily NEP. We preferred this method over using the oxygen optode measurements to estimate seasonal NEP, as MAX-250+ does not present issues related to sensor drift and temporal coverage is higher than the optodes (Meléndez et al. 2020). The final uncertainties in the corrected MAX-250+ oxygen values were ±6 and ±7 μmol kg⁻¹ for La Parguera and Cheeca Rocks, respectively.

Discrete observations

Monitoring at these stations includes laboratory-based, high-precision determinations of potentiometric TA, infrared DIC or spectrophotometric pH. High-quality direct measurements of TA (±1–3 μmol kg⁻¹), DIC (±2 μmol kg⁻¹), and pH (±0.006) were available at each site to fully constrain the carbonate system, and for validation and quality assurance/control processes. Data were available for all seasons. For a complete description of the discrete surveys, the intervals of available data from autonomous and discrete observations refer to Meléndez et al. (2020).

Modeled carbonate parameters

Regional TA–salinity relationships (Lee et al. 2006) tend to break down for nearshore waters subject to freshwater inputs and biological activity; therefore, site-specific relationships must be determined (see Supporting Information Text S1 for further details). The resultant multivariate linear relationship for La Parguera (TA_LP) and Cheeca Rocks (TA_CR) are:

$$\text{TA}_{\text{LP}} = 43.2 \times \text{SSS} - 12.5 \times \text{SST} + 1118.1,$$

$$\text{TA}_{\text{CR}} = 2.2 \times \text{SSS} - 11.8 \times \text{SST} + 2612.7.$$  

(1)

(2)

The modeled uncertainty in TA at La Parguera and Cheeca Rocks were 30 and 47 μmol kg⁻¹, respectively. We used \(p\text{CO}_{2\text{sw}}\) and TA to calculate the \(\Omega_{\text{arag}}\), which had a final uncertainty of 2% (La Parguera) and 4% (Cheeca Rocks), using the CO2SYS (van Heuven et al. 2011) modeling program of Orr et al. (2018). We applied the \(K_1\) and \(K_2\) dissociation constants of Lueker et al. (2000) and \(K_{\text{HSO}_4}\) from Dickson (1990). The solubility constant used to derive \(\Omega_{\text{arag}}\) was from Mucci (1983). Even though the uncertainties in the modeled TA were high, due primarily to calcification and dissolution (Meléndez et al. 2020), the method used here to calculate \(\Omega_{\text{arag}}\) yielded uncertainties that were within the “weather” goal targets (Newton et al. 2015).

Statistical analyses and seasonal changes

We removed high-frequency events, low-amplitude trends, and episodic events before calculating daily averages using a 33-h low-pass filter (Flagg et al. 1976). We define “high-frequency” as those events occurring at 24 h. We seek to decrease the bias that episodic events and random noise can produce, and eliminate signals affecting \(p\text{CO}_{2\text{sw}}\) occurring at sub-daily frequencies such as semidiurnal tides and diurnal heat flux. High-frequency variability affected the seasonal \(p\text{CO}_{2\text{sw}}\) amplitude by 4.5% at Cheeca Rocks, and 0.1% at La Parguera.

Average daily observations were used to construct the annual climatology of the measured and modeled parameters (see Supporting Information Table S2 for a list of the parameters used). The composite year was constructed by binning the data within the representative Julian day. The seasonal variability was computed using the peak-to-peak amplitude and compiled into four distinct seasons: winter (January–March), spring (April–June), summer (July–September), and fall (October–December). Parametric Student t-tests (\(\alpha = 0.05\) significance level) were used to ascertain significance between seasons and differences observed between sites.

Numerical approach: Diagnostic single box mass balance model

To constrain the variability in air–sea exchange, mixing, and biological activity, we assumed an open system and calculated the daily changes in the initial observed values of in situ \(p\text{CO}_{2\text{sw}}\). The carboneic acid system was derived and manipulated using the CO2SYS computer software program (van Heuven et al. 2011) using measured in situ \(p\text{CO}_{2\text{sw}}\) and modeled reef TA (Eqs. 1 and 2).

A 1D carbon mass balance model (Gruber et al. 1998; Shadwick et al. 2011; Fassbender et al. 2016) was used to calculate the net change in surface \(p\text{CO}_{2\text{sw}}\) (\(\delta p\text{CO}_{2\text{sw}}\)) based on the daily partial changes \((\partial t)\) due to gas solubility as a function of temperature and salinity (SOL), air–sea exchange (AIR–SEA EX), mixing processes (HOR MIX), and biology (BIO). The biological term includes water column and benthic organic and inorganic fluxes. Figure 2 shows a schematic diagram of the open carbon box model and the processes assumed to affect NEP and NEC over time within a mixed water parcel close to the buoy. The following equation describes the mass balance of \(\delta p\text{CO}_{2\text{BYS}}\):

$$\frac{\partial p\text{CO}_{2\text{BYS}}}{\partial t} = \frac{\partial p\text{CO}_{2\text{SOL}}}{\partial t} + \frac{\partial p\text{CO}_{2\text{AIR-SEA EX}}}{\partial t} + \frac{\partial p\text{CO}_{2\text{HOR MIX}}}{\partial t},$$

(3)

Thermodynamic variability

The CO2SYS program was used to calculate \(\delta p\text{CO}_{2\text{SOL}}\) using observed daily changes in temperature and salinity. We preferred this method over the temperature-only dependence coefficient (0.0423 °C⁻¹) by Takahashi et al. (1993) because both temperature and salinity create thermodynamic variability in these regions, and the temperature distributions are very different from those in the North Atlantic waters on which this dependence was derived.

Physical transport

The variation attributable to horizontal transport via advection (\(\delta p\text{CO}_{2\text{HOR MIX}}\)) was characterized empirically using daily salinity changes with assumed conservative mixing of the TA
and DIC between the reef and open ocean (Xue et al. 2016). Note that the freshwater contributions from the Amazon River to both the North and South Atlantic oceans do not appear to be sufficient to alter the slope of the TA–salinity relationship (Cai et al. 2010). Therefore, we did not explicitly consider the slope–discharge relationships because according to the linear DIC–salinity and TA–salinity relationships at these sites, the ocean slopes are minimally affected by freshwater inputs and follow the end-member mixing line. The ocean end-member was determined from discrete TA and DIC samples obtained on seasonal cruises (Supporting Information Fig. S2) around the Caribbean and Atlantic regions as follows:

\[
\begin{align*}
\text{TACaribbean} & = 58.8 \times \text{SSS} + 242, \\
\text{DICCaribbean} & = 55.8 \times \text{SSS} + 1.84, \\
\text{TAFlorida Straits} & = 83.8 \times \text{SSS} - 660, \\
\text{DICFlorida Straits} & = 77.1 \times \text{SSS} - 771,
\end{align*}
\]

where \( \pm 5, \pm 12, \pm 3, \) and \( \pm 6 \) are the root mean square errors (RMSEs) of the derived quantity. The slopes of the TA–salinity relationship estimated here were similar to the results of Cai et al. (2010) for the Caribbean and Florida Straits (more details in Supporting Information Text S2). We used discrete TA and DIC measurements from the third Gulf of Mexico Ecosystems and Carbon Cycle cruise along the new 80.6°W transect between Cuba and Florida (Barbero et al. 2019), potentially capturing Gulf Stream water that flows northward before passing to the Cheeca Rocks buoy. Strong linear correlations between TA–salinity and DIC–salinity were observed (\( r^2 > 0.95 \)), supporting the assumption of conservative mixing mechanisms between freshwater and seawater end-members in these regions (Supporting Information Fig. S2). Measurements were taken in August 2017, and those taken at <50 m depth were selected based on the buoyancy frequency (also known as Brunt–Väisälä frequency).

Changes due to vertical mixing were neglected, and the mixed layer (3 m) was assumed to extend to the bottom given the small variations in salinity and temperature observed from Conductivity, Temperature and Depth (CTD) casts performed during the discrete surveys, as well as the shallowness of the sites (Supporting Information Text S3). The seasonal change in salinity due to the mean potential evapotranspiration to precipitation rate was assumed to be small (< 0.1%) and hence neglected. The semidiurnal tides at La Parguera and Cheeca Rocks exhibited amplitudes of approximately 10–20 cm s\(^{-1}\) and oscillated about the mean tidal range (< 0.25 m) of the low-frequency spectrum. The contributions from short-term advection to salinity variability were assumed to be small at both sites.

**Air–sea exchange**

The air–sea CO\(_2\) exchange (\( \partial p_{CO_2} \text{AIR–SEA EX} \)) perturbation is related to DIC changes and air–sea CO\(_2\) flux (\( F_{CO_2} \)). The \( \text{DIC}_{\text{AIR–SEA EX}} \) (\( \mu \text{mol kg}^{-1} \text{ d}^{-1} \)) is estimated via the change in the water column DIC inventory as follows:

\[
\text{DIC}_{\text{AIR–SEA EX}} = \frac{k \times s \times (p_{CO_2^{\text{sw}}} - p_{CO_2^{\text{air}}})}{h \times \rho},
\]

where \( p_{CO_2^{\text{sw}}} - p_{CO_2^{\text{air}}} \) is the difference in atmospheric and seawater CO\(_2\) calculated from the buoy measurements, \( s \) (mol kg\(^{-1}\) atm\(^{-1}\)) is the solubility of CO\(_2\) per unit volume of seawater (Weiss, 1974), \( k \) (m s\(^{-1}\)) is the transfer velocity as a
function of wind speed at 10 m above mean sea level, \( h \) is the water column depth (m), and \( \rho \) is the seawater density (kg m\(^{-3}\)). The transfer velocity–wind speed relationship used was described by Wanninkhof (2014). The net oxygen air–sea flux (\( F_{O_2} \)) was calculated similarly to \( F_{CO_2} \) (see Supporting Information Text S4 for further details). As a convention in this paper, positive \( O_2 \) and \( CO_2 \) fluxes represent transfers from the ocean to the atmosphere.

**Biology**

The biological processes affecting \( pCO_{2sw} \) (\( \partial pCO_{2BIO} \)) were estimated as the residual of the remainder of the other terms in the mass conservation equation (Eq. 3) to close the system. The Revelle factor (\( \beta, \partial ln pCO_{2sw}/\partial ln DIC \)) was used to convert changes in \( \partial pCO_{2BIO} \) to changes in DIC (\( \partial DIC_{BIO} \)) using the relation between \( pCO_{2sw} \) and DIC as defined by Revelle and Suess (1957). Discrete pH and TA measurements were used to calculate the bottle \( pCO_{2sw}, DIC, \) and \( \beta \) values using the CO2SYS program and the same dissociation constants used for the autonomous measurements. The partial change in DIC due to \( pCO_{2sw} \) (\( DIC/pCO_{2sw} \times \beta, \mu mol \ kg^{-1} \ \mu atm^{-1} \)) was linearly related to temperature (Supporting Information Fig. S3). The \( \partial pCO_{2BIO} \) was converted to \( \partial DIC_{BIO} \) (\( \mu mol \ kg^{-1} \)) according to the following relationships for La Parguera and Checca Rocks:

\[
\frac{\partial DIC_{BIO}}{\partial t} \text{LP} (\pm 1) = 102 \times \text{SST} - 2.1,
\]

\[
\frac{\partial DIC_{BIO}}{\partial t} \text{CR} (\pm 4) = 88.7 \times \text{SST} - 1.6,
\]

where \( \pm 1 \) and \( \pm 4 \) are the RMSEs of the derived quantity. In this study, the \( \partial DIC_{BIO} \) was assumed to come from the combined effects of organic matter formation and remineralization together with calcium carbonate precipitation and dissolution:

\[
\frac{\partial DIC_{BIO}}{\partial t} = \frac{\partial DIC_{BIO}}{\partial t} \text{NEP} + \frac{\partial DIC_{BIO}}{\partial t} \text{NEC}.
\]

The changes in oxygen due to biology (\( O_{2BIO} \)) were constrained by removing the effects of solubility using oxygen at saturation \( [O_{2sat}] \) as a function of temperature and salinity (Weiss, 1970). The change in oxygen due to biological production was then calculated as \( [O_{2BIO}] = [O_{2obs}] - [O_{2sat}] \).

**NEP and NEC rates**

The net daily changes in oxygen due to organic production (\( \partial O_{2NEP} \)) are defined as follows:

\[
\frac{\partial O_{2NEP}}{\partial t} = \frac{\partial O_{2BIO}}{\partial t} - \frac{\partial O_{2GAS}}{\partial t},
\]

where \( \partial O_{2BIO} \) is the observed daily change in oxygen concentration (\( O_{2BIO}, \ \text{mmol m}^{-3} \)) due to biological production. The \( \partial O_{2GAS} \) (mmol m\(^{-2}\) d\(^{-1}\)) is the oxygen change due to \( F_{O_2} \) and corrected for bubble flux (more information in Supporting Information Text S4). The changes in \( \partial O_{2NEP} \) were converted to DIC changes (\( \partial DIC_{NEP} \)) using the stoichiometric metabolic quotient (\( Q \)) typically assumed (McGillis et al. 2011; Takegoshi et al. 2016) for coral reef ecosystems (\( Q = 1 \)); that is, for each oxygen molecule consumed, a carbon molecule is produced. An internal analysis was performed using different values of \( Q \), ranging from 0.5 to 1.5 (Bolden et al. 2019), to address potential systematic errors introduced in the metabolic calculations (Supporting Information Text S5 and Fig. S3).

NEC is estimated as follows:

\[
\frac{\partial DIC_{NEC}}{\partial t} = \frac{\partial DIC_{BIO}}{\partial t} - \frac{\partial DIC_{NEP}}{\partial t}.
\]

We subsequently use the terms net autotrophy (net photosynthesis) or net calcification to refer to values of NEC (production – respiration) or NEC (calcification – dissolution) that are greater than zero; similarly, we use the terms net heterotrophy (net respiration) and net dissolution to refer to values of NEP or NEC that are less than zero.

**Uncertainty assessment**

Model errors for the \( pCO_{2sw} \) and DIC mass budget variables, as well as NEP and NEC, were estimated using Monte Carlo simulations (Supporting Information Table S2). Before the Monte Carlo simulations, a Kolmogorov–Smirnov test was performed on each of the model variables to verify that the data were normally distributed. Random normal distributions of each variable were generated using the MATLAB function \( \text{randn} \) (MATLAB 2019a, MathWorks). Sampling repeated 1000 times to establish the final uncertainty, mean, and standard deviation for each simulation. NEC uncertainties included uncertainties from \( pCO_{2sw} \) and NEP. The uncertainty in the gas exchange coefficient was estimated to be 30% (Wanninkhof, 2014).

**Results**

**Seasonal physical and biological drivers of \( pCO_{2sw} \)**

The results from the 1D mass balance model showed that the seasonal amplitudes in \( \partial pCO_{2BIO} \) were less pronounced at La Parguera (25 \( \mu atm \)) than at Cheeca Rocks (100 \( \mu atm \)) because of latitudinal differences in temperature as well as differences in biological composition between the sites (Fig. 3). At Cheeca Rocks, the biological activity was the largest contributor to the observed \( pCO_{2sw} \) variability, with counteracting effects of solubility (temperature and salinity) and physical transport (air–sea \( CO_2 \) gas exchange and horizontal mixing). \( \partial pCO_{2BIO} \) at Cheeca Rocks varied seasonally by 138 \( \mu atm \), with a maximum of 75 \( \mu atm \) in November and a minimum of –63 \( \mu atm \) in April. The values of \( \partial pCO_{2BIO} \) at Cheeca Rocks ranged from –13 \( \mu atm \) in January to +42 \( \mu atm \) in August.
Annual and seasonal net ecosystem metabolism

On an annual basis, both sites were net heterotrophic and net dissolving (Table 1), with occasional periods of net calcification and net autotrophy (Fig. 4). La Parguera and Cheeca Rocks showed similar ($t = -1.0$, df = 22, $p$-value = 0.32) monthly rates of NEP and NEC. The magnitude and direction of the daily NEP and NEC were also similar at both stations (Fig. 4). The net rates of respiration ($\Delta$NEP) and dissolution ($\Delta$NEC) were approximately 1.5 times higher during the summer (July and August) at La Parguera than at Cheeca Rocks. Photosynthesis ($\Delta$NEP) and calcification ($\Delta$NEC) were higher and more variable at Cheeca Rocks on an annual basis. The annual cycles of NEP and NEC at both sites indicate that most of the net organic carbon fixation and calcification occurred from December to May, with higher values from January to March. In July and August, NEP and NEC slightly increased at Cheeca Rocks.

The census-based budgets (gross production – bioerosion) were estimated at both locations (Supporting Information Text S6). Even though the census-based NEC was greater at the Cheeca Rocks due to the high coral biomass (Table 1), ourchemistry-based NEC over a diurnal cycle likely represented processes from a mix of fine-grained CaCO$_3$ sediments, isolated hard and soft corals, and seagrass (Fig. 1). The census-based NEC at Cheeca Rocks is slightly smaller (2.78 kg CaCO$_3$ m$^{-2}$ yr$^{-1}$) if we include abiotic dissolution ($-0.11 \pm 0.12$ kg CaCO$_3$ m$^{-2}$ yr$^{-1}$). The average biologically mediated dissolution of CaCO$_3$ from bioeroding sponges, microbioerosion, parrotfishes, and urchins at Cheeca Rocks was $-1.27$ kg m$^{-2}$ yr$^{-1}$, which would enhance chemical dissolution. This was higher than that at La Parguera, where the total bioerosion was estimated at $-0.51$ kg CaCO$_3$ m$^{-2}$ yr$^{-1}$. At La Parguera the census ($1.0 \pm 1.6$ kg CaCO$_3$ m$^{-2}$ yr$^{-1}$) and chemistry ($-0.68 \pm 0.91$ kg CaCO$_3$ m$^{-2}$ yr$^{-1}$) NEC annual means showed that the fore reef is accreting, albeit minimally, and with high variability (Table 1). The census-based abiotic dissolution at La Parguera is estimated as $-0.24 \pm 0.07$ kg CaCO$_3$ m$^{-2}$ yr$^{-1}$.

The total uncertainties in NEP and NEC calculations were 38% and 46% at La Parguera and 34% and 60% at Cheeca Rocks, respectively (Supporting Information Table S2). These uncertainties are associated with the uncertainties in the gas exchange coefficient and the TA model. The seasonal amplitudes of NEP and NEC were significant ($t = 1.97$, df = 365, $p$-value < 0.05) despite the model uncertainties for both sites. Note that rather than providing values for a certain date, our mass balance approach yielded a climatological assessment of NEP and NEC.

This Modeling caveats and sampling limitations that should be considered include the use of stoichiometric metabolic quotient (Supporting Information Text S7). Even though the magnitude of NEP rate varied based on the stoichiometric metabolic quotient, the seasonal and annual trends did not change (Supporting Information Fig. S4). We assumed the systematic error from the choice of the metabolic quotient to be negligible compared to the large natural variability at these sites. The model was also rerun using oxygen data from an optode

![Fig. 3. Cumulative seasonal changes in surface $pCO_{2sw}$ ($\Delta pCO_{2sw}$; $\mu$atm) are based on contributions from solubility, physical transport, and biological processes for Cheeca Rocks (a) and La Parguera (b). The daily average and model uncertainties (Supporting Information Table S2) in the composite year are represented with solid lines and shaded bounds, respectively. Observations and model results are shown in each panel as observed $pCO_2$ values (OBS; red), effects of temperature and salinity variability (SOL; blue), effects of horizontal mixing (HOR MIX; green), effects of air-sea exchange (AIR-SEA EX; purple), and effects of biological activity (BIO; black).](image-url)
sensor to demonstrate the robust quality of our NEP rates using the post-corrected MAX-250 oxygen measurements (Supporting Information Fig. S6). The observed NEP differences between the two sensors, and the subsequent effects on NEC, were not statistically different ($t_{\text{NEP}} = 0.2$, df $= 361$, $p$-value $= 0.84$).

**Discussion**

**Validation and comparison with other Atlantic–Caribbean reef areas**

There is strong evidence for decreasing carbonate production, coral cover, and biomass of dominant reef-building species throughout the Caribbean and Atlantic regions (Perry et al. 2013). NEC and NEP from this method were validated using census-based budgets and chemistry-based turbulent flux and TA anomaly techniques during short-term surveys at both locations (Supporting Information Figs. S7–S9). Our results agreed with the critical transition zone suggested by Perry et al. (2013), where coral reefs with less than 10% live coral cover typically start to move into a state of net erosion (Supporting Information Fig. S10). The low chemistry-based NEC (Table 1) are not particularly anomalous when compared with the average NEC ranges reported for forereef areas in the Atlantic (0.9–2.7 kg CaCO$_3$ m$^{-2}$ yr$^{-1}$; Vecsei, 2004) and given the variability (>50%) reported for census-based approaches.

We expect that other significant TA and DIC fluxes from organic matter production (Kim and Lee, 2009), anaerobic sediment diagenesis (Yamamoto et al. 2015), CaCO$_3$ sediment calcification/dissolution (Stoltenberg et al. 2021), and nutrient transformation in macroalgae, sediments, and plankton (Wolf-Gladrow et al. 2007) were also integrated into our metabolic measurements. While these assumptions do not affect our calculations, they need to be considered when comparing our results with those of previous studies and for other reef areas. Differences with other studies can be related to several factors, including the method used (e.g., mesocosms, gradient flux, Eulerian, Lagrangian, and census-based), benthic communities, water depth, environmental feedback, hydrodynamics, reef morphology, and spatiotemporal scales. Another possibility is that there is a real decline in NEC in these systems, considering some of these studies are from a decade ago.

**Potential drivers related to benthic and water column processes**

Our method is a powerful tool for determining seasonal and interannual net ecosystem metabolism changes in the field. NEC results suggest that CaCO$_3$ dissolution can play a large role in reef-scale NEC in areas with low % live coral. Both

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**Table 1.** NEP and NEC using census- and chemistry-based approaches. The range is shown for the chemistry-based NEC. The annual mean ($\pm$ uncertainty) is shown for the census-based NEC. Rates are expressed in units of kg CaCO$_3$ m$^{-2}$ yr$^{-1}$ or kg C m$^{-2}$ yr$^{-1}$ using a molecular weight of calcium carbonate (100.09 g mol$^{-1}$) and carbon (12 g mol$^{-1}$) for NEC and NEP, respectively.

| Study                | Location     | NEC kg CaCO$_3$ m$^{-2}$ yr$^{-1}$ | NEP kg C m$^{-2}$ yr$^{-1}$ | Coral/calcifier cover (%) | Reef type/method               |
|---------------------|--------------|----------------------------------|-----------------------------|---------------------------|--------------------------------|
| McGillis et al. (2011) | Puerto Rico  | –0.2 to (–0.15)                  | 10                          | Forereef/enclosure and boundary layer |
| Bates et al. (2010)   | Bermuda      | –0.8 to 3.8                      | –0.5 to 1.0                 | 21                         | Barrier reef and lagoon        |
| Courtney et al. (2016)| Bermuda      | 2.2 to 2.3                       | 26                          | Rim reef, census,* and chemical |
| Muehllehner et al. (2016)| Florida    | –0.3 to 0.6                      | –0.14 to 0.27               | 2–7                        | Patch reefs, Be-7               |
| This study            | Florida      | –3.7 to 1.9                      | –0.28 to 0.11               | 2–7                        | Patch reefs, chemistry          |
| This study            | Florida      | 2.9 ± 0.9                        | 30                          | 4                           | Patch reefs, census†            |
| This study            | Puerto Rico  | –3.1 to 0.8                      | –0.38 to 0.10               | 10–11                      | Fore-reef/lagoon, chemistry     |
| This study            | Puerto Rico  | 1.0 ± 1.6                        | 11                          | Fore-reef slope, chemistry†  |

*Bioerosion is not included.
†Abiotic dissolution is not included.

Fig. 4. Annual composites based on daily averages of modeled (a) NEP (mmol C m$^{-2}$ d$^{-1}$) and (b) NEC (mmol CaCO$_3$ m$^{-2}$ d$^{-1}$) for Cheeca Rocks (red) and La Parguera (black). The daily average and model uncertainties in the composite year are represented with solid lines and shaded bounds, respectively. NEP < 0 is representative of net heterotrophy/respiration, and NEP > 0 indicates net autotrophy/photosynthesis. NEC > 0 indicates net calcification, and NEC < 0 indicates net dissolution.

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sites show net dissolution on an annual basis and a short window (~4 months) of net calcification. This window consistently decreased after 2014 at both sites and ranged throughout the time series from 2 to 3.5 (La Parguera) and 2 to 4.5 (Cheeca Rocks) months (Supporting Information Fig. S11). Potential drivers for this decline include the deterioration of coral communities in the Florida Reef Tract (Lirman et al. 2011; Manzello, 2015; Precht et al. 2016) and at La Parguera (Weil et al. 2009) that have caused poor reef health conditions and facilitated shifts to macroalgal dominance reefs.

At Cheeca Rocks, dissolution rates appeared to increase concomitantly with the seagrass biomass and production, which peaked in summer and was at a minimum in winter (Fourquean et al. 2001; Burdige and Zimmerman, 2002). Microbial decomposition of organic matter can also enhance metabolic dissolution through oxic respiration (Cyronak and Eyre, 2016). Respiration rates at Cheeca Rocks may be fueled by seagrass-derived detritus, seagrass secession, sediment remineralization, and dissolved organic carbon excreted from living leaves, rhizomes, and roots and leached from decomposing tissues (Duarte and Krause-Jensen, 2017). The increases in the biomass and productivity of macroalgae (Collado-Vides et al. 2005) and seagrasses (Lirman and Biber, 2000) from winter to early summer coincide with our highest NEP.

At La Parguera, dissolution is likely generated within the carbonate sediments, coral reef framework, and coral rubble. Shallowness and long residence times at La Parguera site can also increase the accumulation of TA flux from the dissolution of carbonate sediments and coral framework. We hypothesized the mangroves are one of the major organic C sources during the fall in La Parguera. The observed heterotrophic conditions coincided with the wet season, the seasonal decrease in salinity caused by the remote influx of freshwater originating from the Orinoco and Amazon River plumes (Corredor and Morell, 2001), and the frequent occurrence of extensive mats of Sargassum washing up in coastal areas (Wang et al. 2019). We hypothesized that the strong mixing and nutrient supply to the surface in the winter could support positive NEP.

The potential effects of temperature driving increased rates of respiration in the summertime were more evident at La Parguera than at Cheeca Rocks (Supporting Information Fig. S12). At La Parguera, NEP and NEC were negatively correlated with temperature ($r^2 = 0.77$), whereas at Cheeca Rocks correlation with temperature was 0.20. The NEP seasonal cycle at both locations may be driven by either a recent decrease in new nutrients leading to overall lower annual NEP or an increase in organic matter delivery driving overall higher levels of respiration that decrease NEP. Additional work should be focused on understanding other environmental drivers and regional sources of nutrients and organic matter in both locations.

**Relationship between NEC and NEP**

Significant correlations between NEC and NEP have been observed in previous studies in the field on hourly (McMahon
Note that a high correlation coefficient was likely because NEC was partially dependent on NEP in our model (Eq. 13), although $pCO_{2sw}$ and oxygen measurements were independent. To better de-convolve the influence of NEP on NEC, it may be necessary to independently calculate more accurate high-frequency fluxes of TA, which was not possible using our data set.

The threshold NEP rate corresponding to the annual zero NEC (NEC = 0) was estimated from the intercept between NEP and NEC. The intercept at Cheeca Rocks was significantly ($p$-value < 0.05) different than zero at $-2.03 \pm 0.99$ mmol m$^{-2}$ d$^{-1}$. Above this value of annual NEP, we would expect a positive annual NEC at Cheeca Rocks. While at La Parguera the intercept ($-0.03 \pm 0.15$ mmol m$^{-2}$ d$^{-1}$) was not significantly different from zero ($p$-value = 0.86).

**Ecosystem metabolism and $\Omega_{arag}$**

The relationships of NEC with $\Omega_{arag}$ provide a better understanding of the changes in $\Omega_{arag}$ due to the combined effects of the rates of change in NEC and NEP on seasonal timescales (Andersson and Gledhill, 2013). We defined the $x$-intercept (Fig. 5) as the seasonal shifts or threshold points at which NEC was zero. This represents the point at which dissolution exceeded calcification. Figure 6 shows that at Cheeca Rocks the $\Omega_{arag}$ where NEC is zero was 4.00 ± 0.02, which agreed with the results of Muehllehner et al. (2016). Results suggested that $\Omega_{arag}$ was not the primary driver of NEC at La Parguera, while at Cheeca Rocks, the linear correlation between NEC and $\Omega_{arag}$ showed that 30% of the variance can be explained by $\Omega_{arag}$. At La Parguera, the two transitions where NEC was zero occurred at 3.75 in April (spring) and 3.50 in December (fall).

Results suggest that systems such as Cheeca Rocks with a low correlation between NEC and NEP (Fig. 5) and where the organic carbon cycle is relatively balanced (i.e., NEP = 0) could have weaker hysteresis and therefore a significant correlation between NEC and $\Omega_{arag}$. The seasonal hysteresis suggests that long-term observations of nearshore seawater $\Omega_{arag}$ may not be a good predictor of long-term changes in NEC, which has also been suggested by recent diurnal studies (Cyronak et al. 2013; McMahon et al. 2013; Shaw et al. 2015).

**Conclusions**

This study derived seasonal changes in NEC and NEP under naturally variable in situ conditions using a mass balance approach at La Parguera and Cheeca Rocks. The 1D mass balance approach used nearly a decade of long-term and high-frequency $pCO_{2sw}$ and oxygen observations as proxies to decouple NEP and NEC of communities close (< 5 km) to the buoys. Both sites were net heterotrophic and showed net dissolution on an annual scale. The average ($\pm$ standard deviation) seawater surface saturation states that complicates the use of first-order derivations of NEC-$\Omega_{arag}$ (Silverman et al. 2007).
NEC ranged from $11 \pm 5$ and $13 \pm 10 \text{mmol} \text{CaCO}_3 \text{m}^{-2} \text{d}^{-1}$ in winter to $-45 \pm 14$ and $-18 \pm 11 \text{mmol} \text{CaCO}_3 \text{m}^{-2} \text{d}^{-1}$ in summer at La Parguera and Cheeca Rocks, respectively. The average NEC ranged from $11 \pm 5$ and $8 \pm 6 \text{mmol} \text{C} \text{m}^{-2} \text{d}^{-1}$ in winter to $-45 \pm 13$ and $-20 \pm 10 \text{mmol} \text{C} \text{m}^{-2} \text{d}^{-1}$ in summer at La Parguera and Cheeca Rocks, respectively. Results suggest that tropical Caribbean reef ecosystems likely exhibit extended periods of net dissolution of highly soluble carbonate minerals based on the similarities in their environmental characteristics.

Our observations showed that NEC, as diagnosed by dissolved oxygen, is a good predictor of NEC in these systems. As such, it is of the utmost importance for monitoring programs to integrate oxygen sensors, as well as changes in the net ecosystem processes, which often regulate local chemistry conditions. Observations support recent studies that show that $\Omega_{arag}$ probably is not the most informative predictor of long-term NEC outside of a controlled experiment. The 1D model can be tested and applied in other reef regions with similar buoys and auxiliary carbonate data to better understand the ocean acidification conditions over different oceanic settings.

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Conflict of Interest
None declared.

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