Eco-materials for Arsenium and Selenium Removal from Aqueous Solutions

VASILE MINZATU1, CORNELIU MIRCEA DAVIDESCUC, MIHAELA CIÓPECC, PETRU NEGREAC, NARCIS DUTEANUC, ADINA NEGREAC*, MARILENA MOTOC1, ANIKO MANEA2
1Politehnica University of Timisoara, Faculty of Industrial Chemistry and Environmental Engineering, 2 Piata Victoriei, 300006, Timisoara, Romania
2Victor Babes University of Medicine and Pharmacy Timisoara, 2 Piata Eftimie Murgu, 300041 Timisoara, Romania

Environmental engineering have a great importance because is dealing with different fields, most important of them being water resource protection. It is well known that natural waters and especially groundwater contain dissolved substances, most of them unharful for human and animal life, but some of them still potentially harmful. Ground waters containing dissolved arsenic and / or selenium are harmful and not suitable for human an animal life. Selenium represents a human life prerequisite microelement which can become toxic when higher quantities are ingested or accumulated. In comparison arsenic has an acute toxic effect over the human body even when very low quantities are ingested. In both cases the maximum amount were limited by OMS at 10 µg L-1. A new technique was used to produce an ecofriendly composite material by doping graphite with iron oxide, which was used for retention of arsenic and selenium from aqueous solutions. Synthesized material was characterized using SEM, EDX, XRD and DTA. Adsorption capacity and adsorption mechanism were established through equilibrium studies. Maximum adsorption capacity was 400 µg As(V) and 625 µg Se(VI) per gram of adsorbent.

Keywords: adsorption, arsenic, eco-material, selenium, toxicity

Natural groundwater composition dependents on the chemical structure and composition of the rocks in which they are quartered [1-5]. Due to that all natural ground waters can dissolve different substances presented in rocks composition; most of them are not a threat to human or animal life [6]. But in some cases ground waters are contaminated with some substances that can represent a real threat to human and animal life. Arsenic and selenium represent two of these contaminants which can be found into the aquifer rocks, from were can be dissolved frequently under form of inorganic compounds with different oxidation numbers [7]. Natural average concentration of arsenic and selenium in groundwater is approximately 2 ppm, concentration situated under the limit recommended by the World Health Organization (WHO) for drinking water being 10 µg L-1 for both elements (arsenic and selenium) [8, 9].

Selenium represents a vital microelement for human life being involved into the correct operation of the thyroid, because is the active component of one enzyme which is regulating thyroid activity. When the human body have not enough selenium some dysfunctions occur leading at hypothyroidism. When the selenium intromission has an optimum value was observed a decrease in the number of complications due to diabetes such as: neuropathy, retinopathy and cataract. The beneficial effect of selenium is due to the fact that it’s suppressing the insulin deficiency from the human body. Selenium become toxic only when large quantities are ingested or when large quantities are accumulated in time, instead of that arsenic present an acute toxic effect on the human body even when is presented in small quantities [10-12]. Arsenic’s high toxicity at large concentrations it is well known from centuries, but only relatively recent studies proved that long time exposure at smallest concentration have negative health effects. Consumption of contaminated water is the main source through which humans come in contact with arsenic and selenium [13, 14]. WHO recommends selenium daily intake of 6-21 µg per day for newborns and children, until 26 - 35 µg per day for adults, contrary recommended arsenic daily intake is zero. At higher concentration selenium becomes toxic leading at acute gastrointestinal, neurological, respiratory issues [8, 9].

During last decades an impressive and urgent problem was represented by the removal or arsenium from waters and especially from ground waters. In order to solve this problem were developed some techniques such as: precipitation -filtration techniques [15], coagulation-precipitation using as coagulation agents aluminum and iron salts or lime water [15, 16], separation using combined processes such as photocatalysis combined with complexation and filtration [17], aeration, chemical oxidation, oxidation-coagulation or oxidation-precipitation [18], electro-coagulation [19], separation using membranes such us nanofiltration, inverse osmosis, electrodialysis [20], ionic exchange [21] and adsorption [15, 22-26]. Techniques like inverse osmosis and electrodialysis are really expensive and not very effective [15, 27], and other techniques like solvent extraction and bioremediation present a double drawback: high operational costs and not been suitable for the treatment of low amount of waters [15, 28]. It is well known that the arsenium have a higher affinity for iron so for the moment the adsorbents based of iron are the most used during arsenium removal process [29-31]. Usually in water arsenium and selenium are found in anionic form, so the technologies used for their removal are similar, and due to that in present paper was studied their removal. Present study aim to prepare, characterize and test an iron based ecofriendly adsorbent used for elimination of arsenium and selenium from aqueous solutions.

* email: adina.negrea@upt.ro, Phone: +40256404192
During experimental work was prepared a composite material formed from iron oxide particles with dimensions around 500 nm fixed onto the surface of the graphite particles. As precursor for graphite production was used starch which is a soluble and ecofriendly substance [32, 33]. Iron oxide nanoparticles can be obtained by using different methods such as: sol-gel [34], precipitation [35], combustion [36], and by Pechini method [37]. In present study iron oxide nanoparticles were obtained using iron chloride as precursor. The novelty of present study is represented by the preparation of a new adsorbent material based on iron oxide supported on graphite starting form soluble starch and iron chloride. Such material was synthesized by thermal decomposition in protective atmosphere [32, 33, 38].

Experimental part

Materials and methods

Chemicals

Chemical materials used for synthesis of new ecofriendly adsorbent include ferric chloride (FeCl₃, ≥97%) and starch ((C₆H₁₀O₅)n) were purchased from Sigma Aldrich. Arsenical and selenium stock solutions have an initial concentration of 10 g L⁻¹ and were purchased from Merck.

Synthesis methods

Eco friendly materials were synthesized by using as raw materials soluble starch and ferric chloride, in a proportion that would ensure a mass proportion C: Fe equal with 10 : 1 into the final material. Raw materials were dissolved in small quantities of water in order to obtain precursors solutions, which were mixed and heated under continuous stirring for at least 10 min, until a homogeneous paste was obtained. After that the obtained paste was dried into the oven at 75°C. In next stage the dried material was subject of a thermal treatment at 500°C for 4 h in inert atmosphere. Obtained material was designed as sample Am-Fe.

Material characterization

The obtained material was characterized through physico-chemical methods such as:

- scanning electronic microscopy (SEM) and through X-ray dispersion analysis (EDX) using the QUANTA FEG 250 microscope.

- RX diffraction, using the X Rigaku Ultima IV Diffractometer (CuKα = 0.154 nm);

- through differential thermogravimetric analysis using the NETZCH STA 449C thermic scale;

Adsorption process study

To track the adsorbent properties of the synthesized material, studies have been done of As(V) and Se(VI) ion adsorption from aqueous solutions. In order to do this was studied the influence of the initial concentration of As(V) and Se(VI) ions on the absorption capacity of the material and on the degree of their elimination from aqueous solutions.

The concentrations of the arsenic and selenium ions being determined through mass inductively coupled with plasma spectrometry ICP-MS, Aurora M90 BRUKER. Thus there were adsorption studies done for different initial concentrations of As(V) ions (100, 300, 500, 700, 1000, 1500, 2000, 3000 and 4000 µg L⁻¹) and Se (VI) (100, 300, 500, 700, 1000, 1500, 2000, 3000, 3500, 4000, 5000 and 6000 µg L⁻¹), at the temperature of 298 K and for a contact time of 60 min., using the Julabo SW 23 shaker with 200 rot/min.

Starting from the physico-chemical analysis of the synthesized material and based on literature studies, it was found that the presence of iron can positively influence the adsorption of As(V) and Se(VI) from aqueous solutions. The solid-liquid adsorptions take place, in most cases, according to the Langmuir isotherm:

$$\theta = \frac{K_L c}{1 + K_L c},$$

where $K_L$ is the Langmuir constant and $c$ is the concentration of the adsorbed species, its linearization allowing the calculation of the maximal adsorption capacity [39].

Results and discussions

Characterization of the synthesized material

Obtained eco-friendly adsorbent material was characterized by using electronic scanning microscopy (SEM) and X-ray dispersion analysis (EDX), X Ray diffraction, and also thermogravimetric analysis.

The morphology of the synthesized material was highlighted through electronic scanning microscopy SEM (fig. 1) and X-ray dispersion analysis X-EDX (fig. 2).

By analyzing data depicted in figure 1 can observe that the synthesized material is composed of a support structure based on carbon on which are fixed particles of FexOy.

From figure 1a it is observed that the FexOy particles are evenly distributed onto the carbon surface, and also can be observed a relative uniformity of iron oxide particle size. In the detailed picture presented in figure 1b it is observed that the dimensions of these particles have dimensions situated between 500-1600 nm.

Also in order to confirm that the desired material was obtained we recorded the EDX spectra of synthesized Am-Fe sample. From data depicted in figure 2 can observe that the obtained sample contain C, Fe and O atoms, confirming in this way that we obtained the desired material. Simultaneously the production of desired material was confirmed by recoding X Ray diffraction spectra (depicted in fig. 3). By analyzing the spectral lines observed into the recorded spectra and by comparing them with data from ICDD can observe the presence of two different crystalline
forms in synthesized material. Based on that can conclude that we obtain the graphite doped with magnetite particles.

Recorded thermogram obtained when produced material was heated until 1000°C is depicted in figure 4. Based on that can observe that Am - Fe material suffer a small mass change around 100°C, loss associated with water evaporation. By further increase of temperature can observe that the material is stable until around 300°C, when is observed a large mass loss (86.92 %), which is associated with combustion of graphite particles of Am - Fe material. By further increase of temperature can observe that the remaining part of studied material is stable until the maximum temperature was reached. By counting the mass loss during thermogravimetric study can evaluate the ratio C : Fe, and can observe that the value is 11.34 : 1, ratio situated near the desired ratio for studied adsorbent material [40].

Based on data obtained form physico-chemical analysis it can be stated that:
- Soluble starch can be used as precursor for production of graphite support used in adsorbent materials.
- Iron chloride was used as soluble precursor able to modify the surface of graphite support, producing during thermal treatment magnetite particles fixed onto the support surface.
- The synthesized material is formed from two different phases: carbon phase in form of graphite and the iron oxide phase in form of magnetite, which represent a material with good adsorbent properties as is well known from literature.
- Magnetite particles present a particle dimensions ranged between 500 and 1500 nm.
Application of the produced material for arsenic and selenium removal

After synthesis and characterization of desired adsorbent material, were carried out arsenic and selenium adsorption experiments. Experimental data obtained for arsenic adsorption are depicted in figure 5. From these data can observe that the material adsorption capacity have a linear dependence over initial concentration until this concentration is reaching a value of 2000 µg L⁻¹, after this value has reached the adsorption capacity is remaining constant for any further increase of arsenic initial concentration. Also, from this adsorption experiment can conclude that for synthesized material the maximum adsorption capacity was around 410 µg As for each gram of material.

Maximum theoretical adsorption capacity of Am-Fe material can be determined from the linearization of the Langmuir isotherm, which represent the dependence of Ce/q versus Ce. This dependence is described by equation 1. By fitting the experimental data for arsenium adsorption using the Langmuir isotherm [39] was obtained the linear form of the Langmuir isotherm, depicted in figure 6.

\[ \frac{C_e}{q_e} = \frac{1}{K_L q_m} + \frac{C_e}{q_m} \tag{1} \]

By analyzing data presented in figure 6, was obtained the equation associated with linear form of Langmuir isotherm: \( y = 0.0025x + 0.077 \)

Table 1

| Langmuir parameters | K_L, L/µg | q_m, µg/g | R² |
|---------------------|-----------|-----------|----|
| 0.3247              | 400       | 0.999     |

Langmuir parameters associated to the As adsorption process onto Am-Fe are presented in table 1. From data presented in table 1 can observe that the maximum theoretical adsorption capacity have a value near the experimental one.

Similar to the arsenic adsorption was performed the selenium adsorption onto the synthesized Am-Fe material. In first step was followed the influence of selenium initial concentration over the maximum adsorption capacity of Am-Fe material, and the obtained results are presented in figure 7. From analysis of data depicted in figure 7 can observe that the material adsorption capacity is increasing with selenium initial concentration until this concentration is reaching a value of 5000 µg L⁻¹. Further increase of initial concentration, have no influence on the material maximum
adsorption capacity. Also, was determined the maximum adsorption capacity of material which present a value situated between 500 and 650 µg of selenium per each gram of Am-Fe adsorbent material.

Maximum theoretical adsorption capacity can be calculated from equation associated with linear fit of the Langmuir adsorption isotherm, expressed as dependence between Ce/q versus Ce (dependence and linear fit are depicted in fig. 8).

Based on line equation were evaluated the Langmuir parameters associated with selenium adsorption process on Am-Fe (table 2).

Acknowledgement: This work was partially supported by research grant PCD-TC-2017.

References
1. GUO, H., STUBEN, D., BERNER, Z., 2007. 315(1): p. 47-53.
2. YUDOVICH, Y.E., KETRIS, M.P., 2005. 61(3): p. 141-196.
3. BORAH, D., SATOKAWA, S., KATO, S., KOJIMA, T., 2009. 162(2): p. 1269-1277.
4. LUONG, J.H.T., MAJID, E., MALE, K.B., 2007. 1: p. 7 - 14.
5. DUONG, D.D., Adsorption analysis: equilibria and kinetics. Series on chemical engineering. Vol. 2. 1998: London: Imperial College Press.
6. MONDAL, P., MAJUMDER, C.B., MOHANTY, B., 2008. 150(3): p. 695-700.
7. BAIQ, J.A., KAZI, T.G., SHAH, A.O., KANDHRO, G.A., AFRIDI, H.I., KHAN, S., KOLACHI, N.F., 2010. 178: p. 941-948.
8. *** ORGANIZATION, W.H. Selenium in Drinking - water. 2014 [cited 2017 05.05.2017]; Available from: http://www.who.int/water_sanitation_health/dwq/chemicals/selenium.pdf.
9. *** REGISTRY, A.F.T.S.A.D. [cited 2017 05.05.2017]; Available from: https://www.atsdr.cdc.gov/toxprofiles/tp92-c4.pdf.
10. IMRAN, M., AKHTAR, M.S., HASSAN, S., MEHMOOD, A., RUKH, S., KHAN, K.S., KHALID, A., 2015. 27(12): p. 4417 - 4424.
11. TORRES, J., PINTOS, V., GONZALTO, L., DOMINGUEZ, S., KREMER, C., KREMER, E., 2011. 288(1): p. 32-38.
12. PERRONE, D., MONTEIRO, M., NUNES, J., The Chemistry of Selenium. 2015: 3-15.
13. *** AGENCY, U.S.E.P. Environmental Technology Verification Protocol - Protocol for equipment verification testin for arsenic removal. 2003 [cited 2017]; Available from: https://archive.epa.gov/nrmrl/archive-ety/v/web/pdf/039201epadwctr.pdf.
14. CHOONG, T.S.Y., CHUH, T.G., ROBIAH, Y., GREGORY KOY, F.L., AZNI, I., 2007. 217(1): p. 139-166.
15. MOHAN, D., PITTMAN, C.U., 2007. 142: p. 1-53.
16. SONG, S., LOPEZ-VALDIVIESO, A., HERNANDEZ-CAMPOS, D.J., PENG, C., MONROY-FERNANDEZ, M.G., RAZO-SOTO, I., 2006. 40: p. 364-372.
17. MOLINARI, R., ARGURIO, P., 2017. 109: p. 327-336.
18. BORA, A.J., GOGOI, S., BARUAH, G., DUTTA, G., RUKH, S., KHAN, S.K., KOLACHI, N.F., 2010. 178: p. 941-948.
19. GUO, H., STUBEN, D., BERNER, Z., 2007. 315(1): p. 47-53.
20. YUDOVICH, Y.E., KETRIS, M.P., 2005. 61(3): p. 141-196.
21. VITHANAGE, M., HERATH, I., JOSEPH, S., BUNDSCHUH, J., BOLAN, N., OK, Y.S., KIRKHAM, M.B., RINKLEBE, J., 2017. 113: p. 219-230.
22. SUDA, A., MAKINO, T., 2016. 2070: p. 68-75.
23. LATA, S., SAMADDER, S., 2016. 168: p. 387-406.
24. ROHIM, M., MAS HARIS, M.R.H., 2015. 8(2): p. 255-263.
25. ROY, P., MONDAL, N.K., DAS, K., 2014. 2: p. 585-597.
26. QU, D., WANG, J., HOU, D., LUAN, Z., ZHAO, C., 2009. 163: p. 674-679.
27. GHERGIS, T.M., DRUSCHEL, G.K., MCCLESKEY, R.B., HAMERS, R.J., BAINFIELD, J.F., 2001. 35(19): p. 3957-3962.
28. ROBERTS, L.C., HUG, S.J., RUETTIMANN, T., BILLAH, M.M., KHAN, A.W., RAHMAN, M.T., 2004. 38(1): p. 307-315.
29. SAHA, B., BAINS, R., GREENWOOD, F., Separation Science and Technology, 2005. 40: p. 2909-2932.
31. NEGREA, A., LUPA, L., CIOPEC, M., LAZAU, R., MUNTEAN, C., NEGREA, P. 2010. 28(6): p. 467-484.
32. ZHANG, M., PAN, G., ZHAO, D., HE, G., 2011. 159: p. 3509-3514.
33. LIANG, Q., ZHAO, D., 2014. 271: p. 16-23.
34. DURAE, L., MOUTINHO, A., SEABRA, I.J., COSTA, B.F.O., DE SOUSA, H.C., PORTUGAL, A., 2011. 190: p. 548-560.
35. YAZDANI, F., SEDIGH, M., 2016. 184: p. 318-323.
36. BLUTHARDT, C., FINK, C., FLICK, K., HAGEMEYER, A., SCHLICHTER, M., VOLPE, A., 2008. 137: p. 132-143.
37. Gharibshahian, M., Mirzaee, O., Nourbakhsh, M.S., 2017. 425: p. 48-56.
38. Chang, P.R., Yu, J., Ma, X., Anderson, D.P., 2011. 83: p. 640-644.
39. Langmuir, I., 1918. 40(9): p. 1361-1403.
40. Jankovic, B., 2013. 95: p. 621-629.

Manuscript received: 16.09.2018