Applying IUCN Red List criteria to birds at different geographical scales: similarities and differences

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Abstract
Applying IUCN Red List criteria to birds at different geographical scales: similarities and differences. Extinction risk and conservation status of species are assessed at the global scale by the International Union for Conservation of Nature (IUCN). To ensure objectivity, repeatability and traceability, assessments follow a standardized process that uses reliable and verifiable information. Assessments are synthesized according to guidelines, which have recently been adjusted for application at sub–global scales. Nevertheless, species may have several, different or overlapping conservation status. To quantitatively compare assessments from global to sub–national scales, in this study we analyzed 15 assessment lists for 66 game bird species in France. Assessments were made following IUCN guidelines. Overall, our results reveal that (1) assessments at large spatial scales tend to give lower threat status than small–scale assessments; (2) large–scale assessments made it possible to formally verify information whereas smaller–scale assessments usually did not; (3) large–scale assessments are more likely to be based on standardized evidence of reduction in population size and are less exposed to ‘scale–effects’ and ‘edge–effects’; (4) large–scale assessments are also more often based on scientific literature sensu stricto; and (5) sources are more accurately synthesized than Red Lists at small spatial scales. Our results suggest that small–scale Red Lists do not fully match IUCN guidelines and differ significantly in their assessment processes when compared to global standards. The use of subjective and unreliable data in small–scale Red Lists (above all in national and sub–national lists) may jeopardise the original aim of IUCN Red Lists to provide comprehensive and scientifically rigorous information, and could thus compromise the credibility and prestige of IUCN Red Lists in the eyes of researchers, the general public, and other stakeholders.

Key words: Biodiversity assessment, Game bird species, Conservation status, Information–based management, IUCN Red Lists, Regional assessment

Resumen
Aplicación de los criterios de la Lista Roja de la Unión Internacional para la Conservación de la Naturaleza en diferentes escalas geográficas a las aves. La Unión Internacional para la Conservación de la Naturaleza (IUCN) se encarga de evaluar a escala mundial el riesgo de extinción y el estado de conservación de las especies. Para garantizar su objetividad, repetibilidad y trazabilidad, en las evaluaciones se sigue un proceso estandarizado que hace uso de información fiable y verificable. Asimismo, las evaluaciones se sintetizan de acuerdo con determinadas directrices, que se han ajustado recientemente para su aplicación a escalas inferiores. No obstante, la misma especie puede clasificarse en varios estados de conservación, distintos o superpuestos. Para comparar cuantitativamente las evaluaciones de escala mundial a escala subnacional, analizamos 15 listas de evaluación relativas a 66 especies de aves cinegéticas en Francia; según se había declarado, dichas evaluaciones se realizaron en consonancia con las directrices de la UICN. En general, nuestros resultados ponen de manifiesto que (1) las evaluaciones a gran escala espacial tienden a dar como resultado estados de peligro inferiores que las de pequeña escala; (2) las evaluaciones a gran escala permitieron comprobar de forma oficial la información en que se basan, mientras que las evaluaciones a menor escala no; (3) las evaluaciones a gran escala son más propensas a basarse en pruebas estandarizadas de una reducción significativa.
del tamaño de la población y a estar menos expuestas a efectos de 'escala' o de 'borde de distribución'; (4) las evaluaciones a gran escala también se basan más frecuentemente en publicaciones científicas en sentido estricto; y (5) las fuentes se sintetizan con mayor exactitud en comparación con las Listas Rojas a escalas espaciales pequeñas. Por lo tanto, nuestros resultados sugieren que las Listas Rojas a pequeña escala no coinciden plenamente con las directrices de la UICN y que difieren de forma significativa con respecto a sus procesos de evaluación en comparación con los estándares mundiales. El uso de información subjetiva y poco fiable en las Listas Rojas en pequeña escala (sobre todo en las listas nacionales y subnacionales) puede poner en peligro el objetivo original de las Listas Rojas de la UICN de proporcionar información completa y científicamente rigurosa y, por lo tanto, podría comprometer la credibilidad y el prestigio de las Listas Rojas de la UICN a los ojos de los investigadores, del público en general y de otras partes interesadas.

Palabras clave: Evaluación de la biodiversidad, Especies de aves cinegéticas, Estado de conservación, Gestión basada en información, Listas Rojas de la IUCN, Evaluación regional

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Introduction

Evidence-based wildlife management (Sutherland et al., 2004) requires reliable information on, above all, the conservation status and the extinction risk of species. The most widely recognized assessment of the conservation status of species is the Red List of Threatened Species, established by the International Union for Conservation of Nature (IUCN) (de Grammont and Cuarón, 2006; Rodrigues et al., 2006; Szabo et al., 2012; Maes et al., 2015). The great value of the IUCN Red Lists is derived from their original aim to represent a comprehensive source of scientifically rigorous information (Rodrigues et al., 2006). IUCN assessments have to be objective, transparent, repeatable and traceable (Fitzpatrick et al., 2007; Miller et al., 2007). To this aim, Red Lists are derived from assessments that use data published in a searchable format (Rodrigues et al., 2006). Nevertheless, assessments also routinely use expert knowledge (McBride et al., 2012), so that guidelines for the reliable integration of such knowledge are under development (McCarthy et al., 2004; McBride et al., 2012; Drescher et al., 2013; Drolet et al., 2015). Publication of both data and assessments is now consolidated at global scale through to the on-line searchable IUCN databases accessible via the Internet at http://www.iucnredlist.org. Assessments are constructed using explicitly defined categories and quantitative criteria that are applicable and valid at global scales (Akgakaya et al., 2000; IUCN, 2001). Over the past two decades, these criteria and categories have been revised, thresholds have been adjusted and new categories created (Gärdenfors, 2001; IUCN, 2013). The robustness of assessments has been consolidated through the standardization of data-driven procedures and the use of objective criteria that no longer depend on approaches that entail risks of subjectivity (e.g. threat categorizations based directly on expert opinions) (Mace and Lande, 1991; Rodrigues et al., 2006). Nevertheless, some of the issues that still need to be resolved have been underlined by, for instance, the application of IUCN criteria at sub–global scales (Gärdenfors, 2001; Mace et al., 2008). Hereafter we use ‘sub–global’ as a synonym of ‘regional’ sensu lato to avoid potential confusions with political districts (regions sensu stricto), that in several countries such as France correspond to sub–national administrative territories.

Given that the majority of conservation actions take place at sub–global scales, and that the most influential institutions working on conservation legislation and action are national and regional governments, the concern for the spatial sub–structuring of the threat status of species is increasing (Gärdenfors, 2001; Gärdenfors et al., 2001; Miller et al., 2007). The regional concept (sensu lato) implies a geographically-defined sub–global area, which could be a continent, a country, a state or a province (IUCN, 2012). Guidelines for the application of IUCN Red List criteria at regional levels were published (Gärdenfors, 2001; IUCN, 2012) and assessment of the conservation status of species at sub–global scales were developed (Miller et al., 2007; Azam et al., 2016). These updated guidelines represent the standardized processes that must be applied (without deviation or modification) if regional Red List authorities wish to state that their assessments follow the IUCN system (IUCN, 2012: p. 3). Nevertheless, a risk of subjectivity in the regional adjustment process has been identified, along with the need for more complementary information for identifying national priorities and responsibilities with regards to species’ conservation (Keller and Bollmann, 2004; Rodríguez et al., 2004; Keller et al., 2005).

Despite efforts to create objective processes for assessing species’ extinction risks at sub–global scales, some problems still persist (Gärdenfors, 2001; Martín, 2009; Seoane et al., 2011). Natural scarcity or rarity at local scales may result in the overestimation of threat levels and so the ecological bases of their rarity should be taken much more into account (Martín, 2009; Seoane et al., 2011). Moreover, the second step of the IUCN regional guidelines, which consists of adapting categories according to vaguely formulated terms such as the level of contact with neighbouring populations (Keller et al., 2005: p. 1828), leaves room for interpretation by assessors and a degree of subjectivity (Eaton et al., 2005; Keller et al., 2005). This step is particularly important when assessing very mobile species, such as birds at small landlocked sites (Keller and Bollmann, 2004; Keller et al., 2005) since it requires a lot of accurate data and knowledge on species that is not always available and can be difficult to obtain (Keller and Bollmann, 2004; Eaton et al., 2005; Keller et al., 2005). Indeed, available data for regional assessments are sometimes limited to local populations, and obtaining data for cross–boundary populations is often difficult (Keller et al., 2005). Conversely, in some cases available data may be accurate for large–scale assessments but unsuitable or unreliable for local scales due to limits imposed by their resolution (Hurlbert and Jetz, 2007) and may be negatively affected by confounding methodological factors such as missing data or low number of counts per site (Atkinson et al., 2006).

In addition to these yet unresolved issues, recent concerns about threatened species have led to an increase of regional and global Red Lists, sometimes reflecting different conservation statuses of the same species at different scales. France is a particularly interesting example (Azam et al., 2016), as the conservation status of its birds has been characterized at the global level by IUCN and Birdlife International (IUCN, 2015b), at European level by Birdlife International (BirdLife International, 2015), at the national level by the French Committee for IUCN (UICN France et al., 2011), and at the sub–national level (French regions) by local organizations (Filitti and Vincent–Martin, 2013; LPO Alsace, 2014). As a result, each bird species may be classified under five different conservation statuses in France (detailed below). Thus, in light of the increasing number of Red Lists referring to the same species, the simple question ‘what is the conservation status of the considered bird species?’ has become far more complex for all concerned. The choice to use one or other of these different classification-
tions according to the particular situation may rest on the understanding of their respective characteristics and limits. A quantitative comparative analysis of their characteristics is thus required, but to our knowledge is not yet available.

In this study, we used a set of game bird species as a case study to compare several Red Lists (from sub–national to global scale) that classify the conservation status of birds in France. Previous studies in other countries have compared global and regional Red Lists or analyzed the regional assessment of certain taxa. Their results highlight the fact that regional lists tend to lead up to higher threat statuses than the global IUCN Red Lists due, for instance, to the scale–dependent chance of meeting Red List criteria (hereafter referred to as ‘scale effect’) and to the ‘edge effect’ of small–scale assessments of the conservation status of species (Keller et al., 2005; Milner–Gulland et al., 2006; Brito et al., 2010). Theoretically, different assessments should agree if they use (for instance, for endemic species) a common methodology, identical species and the same information. Nevertheless, in light of the results from other countries (Milner–Gulland et al., 2006), we expect that in this study the assessments of the status of birds (including numerous non–endemic species) at smaller scales would result in higher threat statuses compared to larger scale assessments.

As mentioned above, the variability in conservation status between lists at different scales may be due to ‘scale–’ or ‘edge–effects’, in particular in regional assessments, when criterion D of the IUCN is used, which considers small population size (Eaton et al., 2005; Keller et al., 2005). To evaluate whether the variability in conservation statuses between lists at different scales in France is due to the ‘population size effect’, we compared the proportions of criteria used in assessments. On the basis of the associations reported in previous studies (Eaton et al., 2005; Keller et al., 2005), we expected that a higher proportion of species would be classified as threatened based on criterion D at small–scale as opposed to larger scale assessments.

The differences in status due to the scale of the assessment could also be associated with disparities in the type of information used to evaluate species’ conservation status (de Grammont and Cuárron, 2006). One hypothesis suggests that some lists and evaluations may depend primarily on data from grey literature, which conflicts with the comprehensive, scientifically rigorous and transparent nature of Red Lists (Mrosovsky and Godfrey, 2008). To evaluate whether variability in conservation status between different lists is linked to the differences in the type of information that were synthesized, the proportions of categories of information used in French Red Lists were compared. On the basis of the reported greater likelihood of subjectivity in small–scale assessments (Eaton et al., 2005; Keller et al., 2005), we predicted that grey literature would play a more important role in small–scale Red Lists than in large–scale assessments.

The distinction between scientific and grey literature has been widely debated (Schöpfel, 2006; Mrosovsky and Godfrey, 2008) and is detailed below in the methods section. Among other characteristics, grey literature is usually less available, less reliable and more incomplete than scientific literature (Conn et al., 2003; Schöpfel, 2006; Mrosovsky and Godfrey, 2008). Moreover, literature unavailability has been proposed as a factor that might be associated with unreliable citations (Todd and Ladle, 2008). Thus, according to this hypothesis, we predicted that grey literature might be more associated with cases of no supported citations than scientific literature.

Material and methods

Study species

Highly mobile species may be more affected by the challenges posed by regional adaptations to the IUCN system, particularly when applied at small geographical scales and when data on across–boundary population dynamics are required (Akçakaya et al., 2000; Keller and Bollmann, 2004; Keller et al., 2005; IUCN, 2012). To ensure comparability, analyses should be based on overlapping sets of species; as well, data sets should contain well–studied species in order to make quantitative analyses possible. Many birds are widely distributed mobile species that represent a well–studied taxonomic group (van Jaarsveld et al., 1998; Butchart et al., 2004; Fazey et al., 2005). However, it is a vast taxonomic group and so we chose to analyse a subset for potential heterogeneity between Red Lists. Among birds, game species may be the object of multiple and additive conservation actions, such as monitoring programs conducted by wildlife recreationists with a variety of different motivations (Cooper et al., 2015) and hence these species may be better studied than others. Thus, we focused our analyses on 66 game bird species in France (table 1).

IUCN–type Red Lists

France is a biogeographically diverse country crossed by many migratory flyways and there are several Red Lists for birds to assess their conservation status. To ensure comparability, in this study we only analyzed lists based on the IUCN classification system, as these lists are expected to follow the standardized processes detailed in the IUCN guidelines (IUCN, 2001, 2012). The global IUCN Red List was updated during the development of this study at the end of 2015. This allowed us to verify the potential effects of this update on the potential scale–dependent heterogeneity in Red Lists. As a result, we compared 15 lists in this study (table 2): two versions (a pre–update version from October 2015 and an updated version from December 2015) of the IUCN global Red List of Birds (IUCN, 2015a, 2015b); the European Red Lists of Birds (BirdLife International, 2015), available at regional levels for (i) geographical Europe (Europe) and (ii) the member states of the European Union in 2012 (EU27); the national (French) Red List of Birds (IUCN France et al., 2011); and 10 sub–national
Red Lists of Birds for Île-de-France (IDF) (Birard et al., 2012), Limousin (Roger and Lagarde, 2015), Pays de la Loire (PaysLoire) (Marchadour et al., 2014), Midi-Pyrénées (MidiPyr) (Fremaux, 2015), Provence–Alpes–Côte-d’Azur (PACA) (Flitti and Vincent–Martin, 2013), Languedoc–Roussillon (LangRou) (Meridionalis, 2015), Centre (Nature Centre, 2013), Alsace (LPO Alsace, 2014), Bretagne (Bretagne Environnement, 2015) and Bourgogne (Bourg) (Abel et al., 2015). All these assessments declared to have followed the IUCN system; the considered sub–national lists were approved and labelled by the IUCN French committee, a process designed to guarantee that sub–national Red Lists follow IUCN guidelines (IUCN France, 2011; http://iucn.fr/etat-des-lieux-listes-rouges-regionales/).

Compiled data

Conservation status and criteria

For all the species on each Red List, we compiled the conservation status (LC, least concern; NT, near threatened; VU, vulnerable; EN, endangered; CR, critically endangered; RE, regionally extinct) and, whenever possible, the criteria underlying it (A, population reduction; B, geographic range; C, small population size and decline; D, very small or restricted regional population) (IUCN, 2001, 2012). Criterion E (based on quantitative analyses that estimate the probability of extinction; IUCN, 2012) was not present in our sample.

In a few cases (6.56%), the species were qualified as Threatened based on multiple criteria. In such cases, we used the first used criterion according to the classification E > A > B > C > D because small regional populations or local rarities do not necessarily imply a high risk of extinction (Harnik et al., 2012) while criterion A deals with species that are at risk because of a steep rate of decline (Collen et al., 2016).

Bibliographical categories

The references cited in the Red Lists were compiled and classified in four categories on the basis of the following definitions. The first category (A) is ‘scientific literature’ sensu stricto, that is, work published in scientific journals that is indexed in scientific data sources (Björk et al., 2010) and peer reviewed (Steven et al., 2011). Scientific literature meets methodological standards (Conn et al., 2003), is easily available (Pyšek et al., 2008) and represents, notwithstanding certain flaws, a widely accepted strategy for ensuring quality control in scientific research (Ferreira et al., 2015). The second category (B) includes ‘referenced books’, in particular, books identified by an International Standard Book Number (ISBN) and referenced academic publications, such as PhD theses. These data sources may be accessible yet allow some freedom from the peer–review process, and thus they summarize science from a personal perspective to present ideas in a liberating manner (McWilliams and Bauchinger, 2012). The third category (C) is ‘grey literature’ that includes publications that are not peer–reviewed (Conn et al., 2003) and articles that appear in non–indexed journals. Such articles are difficult to identify and to access through classical routes and often lack robust methodology and traceability (Corlett, 2011; Friess and Webb, 2011). Finally, the fourth category (D), ‘expert opinion’, includes estimates based on empirical knowledge or even field experience that is not to be found even in grey literature.

Citation categories

We compiled and classified the way in which data, citations and sources were included in the IUCN–based assessments First, we analysed whether it was possible to link the detailed information in assessments to citations and sources. Next, the reliability of the cited information was classified through consensus between the two authors (MC and MS) into one of four categories as defined in Todd et al. (2010): (1) There is ‘Clear support’, when the cited article provides unequivocal support of the assertion via either statements in the text or the data presented. (2) ‘Ambiguous’, when the material (either text or data) in the cited article has been interpreted one way, but could also be interpreted in other ways, including the opposite. The assertion in the primary article is supported by a portion of the cited article, but that portion runs contrary to the overall thrust of the cited article. The assertion includes two or more components, but the cited article only supports one of them. (3) ‘No support’, when the cited article does not in any way substantiate the assertion via either statements in the text or the data presented. The cited article may even contradict the assertion in the primary article. (4) ‘Empty citation’, when the cited article simply cites other articles that support the assertion made in the primary article. Citing a review article is acceptable if the support for the assertion is, for example, a new insight or opinion offered by the author(s) of the review.

As in Todd et al. (2010), if the cited article was classified as ‘empty citation’ plus ‘no support’, ‘no support’ took precedence. If the cited article was classified as ‘empty citation’ plus ‘ambigous’, ‘ambigous’ took precedence. Another citation category (‘unverifiable’) was created for expert opinions and for cases in which the lack of published or available documents make the assessment of the links to the information source impossible.

Statistical analysis

We used Fisher’s tests (Millot, 2011) to test if the proportion of threatened categories was significantly different between Red Lists for the overlapping sets of species, taking into account the small sample size in some tests and the need for standardized analyses for comparisons. Additionally, the odds ratio (ranging from 0 to infinity) in bilateral tests on contingency tables was used to analyze the direction of detected differences (Millot, 2011). The further away the odds ratio was from 1 towards infinity, the more the first list in the test was characterized by the considered factor. The more the second list in the test was characterized by the considered factor, the closer the odds ratio was to 0.
Table 1. Bird species and conservation status according to IUCN–based Red Lists at different scales. Global (IUCNo, IUCN version from October 2015; IUCNd, IUCN version from December 2015). European (Eur, geographical Europe; EU27, European politico–economic Union). National (UICN–F, French Committee of the International Union for Conservation of Nature). Sub–national (SN: IDF, Île–de–France; Ce, Centre; Al, Alsace; PL, Pays de Loire; Br, Bourgogne; LaR, Languedoc–Roussillon; MiP, Midi–Pyrénées; PACA, Provence–Alpes–Côte–d’Azur; Li, Limousin. Conservation status (LC, least concern; NT, near threatened; VU, vulnerable; EN, endangered; CR, critically endangered; RE, regionally extinct). DD, data deficient; NA, not applicable; NE, not evaluated.

Tabla 1. Especies de aves y estado de conservación según las Listas Rojas basadas en el método UICN a distintas escalas: Global (IUCNo, versión IUCN de octubre de 2015; IUCNd, versión IUCN de diciembre de 2015). Europea (Eur: Lista Roja europea para Europa geográfica; EU27, Lista Roja para la unión politico–económica europea. Nacional (UICN–F, Lista Roja nacional de Francia). Subnacional (SN: IDF, Isla de Francia; Ce, Centro; Al, Alsacia; PL, Pais de Loira; Br, Bretaña; Bo, Borgoña; LaR, Languedoc–Rosellón; MiP, Mediodía–Pirineos; PACA, Provenza–Alpes–Costa Azul; Li, Lemosín). Estatus de conservación: LC, preocupación menor; NT, casi amenazada; VU, vulnerable; EN, en peligro; CR, en peligro crítico; RE, extinto a nivel regional. DD, datos insuficientes; NA, no aplicable; NE, no evaluada.

| Scientific name   | IUCNo | IUCNd | Eur | EU 27 | IDF | Ce | PL | Br | Bo | LaR | MiP | PACA |
|-------------------|-------|-------|-----|-------|-----|----|----|----|----|-----|-----|------|
| Corvus corone     | LC    | LC    | LC  | LC    | LC  | LC | LC | LC | LC | LC  | LC  | LC   |
| Garrulus glandarius | LC    | LC    | LC  | LC    | LC  | LC | LC | LC | LC | LC  | LC  | LC   |
| Pica pica         | LC    | LC    | LC  | LC    | LC  | LC | LC | LC | LC | LC  | LC  | LC   |
| Corvus frugilegus | LC    | LC    | LC  | LC    | LC  | LC | LC | LC | LC | LC  | LC  | LC   |
| Sturnus vulgaris  | LC    | LC    | LC  | LC    | LC  | LC | LC | LC | LC | LC  | LC  | LC   |
| Anas platyrhynchos | LC   | LC    | LC  | LC    | LC  | LC | LC | LC | LC | LC  | LC  | LC   |
| Anas strepera     | LC    | LC    | LC  | LC    | LC  | NA | EN | NT | CR | NT  | CR  | CR   |
| Anas clypeata     | LC    | LC    | LC  | LC    | LC  | CR | EN | NA | LC | EN  | CR  | DD   |
| Anas acuta        | LC    | LC    | LC  | LC    | LC  | CR | EN | NA | LC | EN  | CR  | EN   |
| Anas penelope     | LC    | LC    | LC  | LC    | LC  | CR | EN | NA | LC | EN  | CR  | DD   |
| Anas crecca       | LC    | LC    | LC  | LC    | LC  | CR | EN | CR | CR | CR  | NA  | CR   |
| Anas querquedula | LC    | LC    | LC  | LC    | VU  | CR | CR | NA | VU | CR  | CR  | DD   |
| Aythya ferina     | LC    | VU    | VU  | VU    | VU  | EN | NT | CR | CR | VU  | EN  | CR   |
| Aythya marila     | LC    | VU    | VU  | VU    | VU  | NT | EN | NA | NA | NA  | NA   |
| Aythya fuligula   | LC    | LC    | LC  | LC    | NT  | NT | VU | VU | NT | CR  | VU  | EN   |
| Bucephala clangula | LC   | LC    | LC  | LC    | NA  | NA | NA | NA | NA | NA  | NA   |
| Netta rufina      | LC    | LC    | LC  | LC    | LC  | VU | VU | VU | NT | VU  | VU   |
| Melanitta fusca   | EN    | VU    | VU  | VU    | VU  | EN | NA | NA | NA | NA  | NA   |
| Melanitta nigra   | LC    | LC    | LC  | LC    | LC  | VU | LC | LC | LC | LC  | LC  | LC   |
| Somateria mollissima | LC | NT    | VU  | EN    | CR  | CR | CR | CR | CR | CR  | CR   |
| Clangula hyemalis | VU    | VU    | VU  | VU    | VU  | NA | NA | NA | NA | NA  | NA   |
| Anser anser       | LC    | LC    | LC  | LC    | LC  | VU | NA | EN | NA | EN  | NA   |
| Anser fabalis     | LC    | LC    | LC  | LC    | LC  | VU | NA | NA | NA | NA  | NA   |
| Anser albinfon    | LC    | LC    | LC  | LC    | LC  | NA | NA | NA | NA | NA  | NA   |
| Gallinago gallinago | LC  | LC    | LC  | LC    | LC  | EN | RE | CR | RE | CR  | CR  | RE   |
| Lymnocryptes minimus | LC  | LC    | LC  | LC    | LC  | NA | NA | NA | NA | NA  | NA   |
| Vanellus vanellus | LC    | NT    | VU  | VU    | VU  | VU | VU | VU | VU | VU  | EN  | EN   |
| Pluvialis apricaria | LC  | LC    | LC  | LC    | LC  | LC | LC | LC | LC | LC  | LC  | LC   |
| Scientific name               | IUCN No | IUCN d | Eur | EU 27 | UICN-F | IDF | Ce | Pl | Br | Bo | LR | MiP | PAPA | Li |
|-------------------------------|---------|--------|-----|-------|--------|-----|----|----|----|----|----|-----|------|----|
| Pluvialis squatarola          | LC      | LC     | LC  | LC    | LC     | LC  |    |    |    |    |    |     |      |     |
| Haematopus ostralegus         | LC      | NT     | VU  | VU    | LC     |     | EN | VU | EN | EN | NA |     |      |     |
| Numenius arquata              | NT      | NT     | VU  | VU    | NA     | CR  | EN | EN | VU | CR | CR | CR  |      |     |
| Numenius phaeopus             | LC      | LC     | LC  | LC    | LC     | LC  |    |    |    |    |    |     |      |     |
| Limosa limosa                 | NT      | NT     | VU  | VU    | RE     | NA  | VU | RE |    |    |    |    |      |     |
| Limosa lapponica              | LC      | NT     | LC  | LC    | LC     | LC  |    |    |    |    |    |     |      |     |
| Tringa totanus                | LC      | LC     | LC  | VU    | LC     | RE  | LC | EN | EN | EN | NA |     |      |     |
| Tringa nebularia              | LC      | LC     | LC  | LC    | LC     | DD  | EN |    |    |    |    |     |      |     |
| Tringa erythropus             | LC      | LC     | LC  | NT    | NA     | DD  | EN |    |    |    |    |    |      |     |
| Philomachus pugnax            | LC      | LC     | EN  | NT    | NA     |     |    |    |    |    |    |    |      |     |
| Calidris canutus              | LC      | NT     | LC  | LC    | NT     | LC  |    |    |    |    |    |     |      |     |
| Fulica atra                   | LC      | LC     | NT  | LC    | LC     | LC  | LC | LC | VU | LC | EN |     |      |     |
| Gallinula chloropus           | LC      | LC     | LC  | LC    | DD     | VU  | VU | VU | DD | EN | LC | EN   |      |     |
| Rallus aquaticus              | LC      | LC     | LC  | DD    | VU     | VU  | VU | DD | EN | LC | EN | EN   |      |     |
| Scolopax rusticola            | LC      | LC     | LC  | LC    | NT     | NT  | LC | NT | LC | VU | DD | NT   | DD   |     |
| Columba palumbus              | LC      | LC     | LC  | LC    | LC     | LC  | LC | LC | LC | LC | LC |     |      |     |
| Columba oenas                  | LC      | LC     | LC  | LC    | LC     | LC  | LC | LC | DD | VU | VU | VU   | VU   |     |
| Columba livia                 | LC      | LC     | LC  | LC    | LC     | LC  | LC | LC | LC | DD | VU | VU   | VU   |     |
| Streptopelia turtur           | LC      | VU     | VU  | NT    | LC     | NT  | LC | NT | LC | VU | LC | LC   | VU   |     |
| Streptopelia decaocto         | LC      | LC     | LC  | LC    | NT     | NT  | LC | NT | LC | VU | LC | LC   |      |     |
| Turdus viscivorus             | LC      | LC     | LC  | LC    | LC     | LC  | LC | LC | LC | LC | LC | LC   |      |     |
| Turdus philomelos             | LC      | LC     | LC  | LC    | LC     | LC  | LC | LC | LC | LC | LC | LC   |      |     |
| Turdus pilaris                | LC      | LC     | LC  | VU    | LC     | NA  | NA | VU | DD | EN | VU | CR   | LC   |     |
| Turdus iliacus                | LC      | NT     | NT  | VU    | LC     | NA  | DD |    |    |    |    |     |      |     |
| Turdus merula                 | LC      | LC     | LC  | LC    | LC     | LC  | LC | LC | LC | LC | LC | LC   |      |     |
| Coturnix coturnix             | LC      | LC     | LC  | LC    | NT     | LC  | NT | LC | DD | NT | VU | NT   |      |     |
| Alauda arvensis               | LC      | LC     | LC  | LC    | LC     | NT  | NT | NT | LC | NT | LC | LC   |      |     |
| Tetrao urogallus              | LC      | LC     | LC  | LC    | LC     | VU  | CR |    |    |    |    |     |      |     |
| Tetrao tetrix                 | LC      | LC     | LC  | LC    | LC     | RE  |    |    |    |    |    |    |      |     |
| Lagopus muta                  | LC      | LC     | NT  | VU    | LC     |     |    |    |    |    |    |    |      |     |
| Tetrastes bonasia             | LC      | LC     | LC  | LC    | VU     | CR  |    |    |    |    |    |    |      |     |
| Alectoris graeca              | NT      | NT     | NT  | VU    | NT     | RE  |    |    |    |    |    |    |      |     |
| Perdix perdix                 | LC      | LC     | LC  | LC    | LC     | NT  | EN | NE | DD | DD | CR | RE   | NA   | DD |
| Alectoris rufa                | LC      | LC     | LC  | LC    | LC     | DD  | LC | NA | NE | DD | DD | DD   |      |     |
| Phasianus colchicus           | LC      | LC     | LC  | LC    | LC     | NE  | LC | NE | DD | LC | NA | LC   | DD   |     |
| Symatricus reevesii           | VU      | VU     | NA  | NA    | NA     | NA  |    |    |    |    |    |    |      |     |
| Callipepla californica        | LC      | LC     | NA  | NA    | NA     |     |    |    |    |    |    |    |      |     |
| Colinus virginianus           | NT      | NT     | NA  | NA    | NA     |     |    |    |    |    |    |    |      |     |
Table 2. Characteristics of the 15 compared lists of birds: Y, year; Sc, scale (G, global; E, European; N, national; Sn, sub-national); Gs, game species; Ca, IUCN categories; Cr, detailed IUCN criteria; Sr, sources; As, assessment–sources link; T, total number of sources useful for assessment; Pr, proportion of references obtained among all sources useful for assessment; \(^1\) proportion of references obtained among cited and referenced references used in the status assessment. (For other abbreviations see table 1).

| IUCN–based Red Lists of Birds | Y | Sc | Gs | Ca | Cr | Sr | As | T | Pr |
|-------------------------------|---|----|----|----|----|----|----|---|----|
| IUCN (version from October 2015) | 2015 | G | 66 | ✓ | ✓ | ✓ | ✓ | ✓ | ✓ |
| IUCN (version from December 2015) | 2015 | G | 66 | ✓ | ✓ | ✓ | ✓ | ✓ | ✓ |
| Europe (geographical Europe, EU27) | 2015 | E | 63 | ✓ | ✓ | ✓ | ✓ | ✓ | ✓ |
| UICN French Committee | 2011 | N | 66 | ✓ | ✓ | ✓ | ✓ | ✓ | ✓ |
| Bourgogne | 2015 | Sn | 35 | ✓ | ✓ | ✓ | ✓ | ✓ | ✓ |
| Limousin | 2015 | Sn | 59 | ✓ | ✓ | ✓ | ✓ | ✓ | ✓ |
| IDF | 2012 | Sn | 34 | ✓ | ✓ | ✓ | ✓ | ✓ | ✓ |
| Alsace | 2014 | Sn | 40 | ✓ | ✓ | ✓ | ✓ | ✓ | ✓ |
| Centre | 2013 | Sn | 38 | ✓ | ✓ | ✓ | ✓ | ✓ | ✓ |
| Pays de Loire | 2013 | Sn | 40 | ✓ | ✓ | ✓ | ✓ | ✓ | ✓ |
| Languedoc–Roussillon | 2015 | Sn | 40 | ✓ | ✓ | ✓ | ✓ | ✓ | ✓ |
| Midi–Pyrénées | 2015 | Sn | 17 | ✓ | ✓ | ✓ | ✓ | ✓ | ✓ |
| PACA | 2013 | Sn | 40 | ✓ | ✓ | ✓ | ✓ | ✓ | ✓ |
| Bretagne | 2015 | Sn | 51 | ✓ | ✓ | ✓ | ✓ | ✓ | ✓ |

Using Fisher’s tests and based on the odds ratio we also examined (i) the proportion of added criteria for threatened categories, (ii) the proportion of bibliographical categories and (iii) the proportion of citation categories in assessments among the different scales of Red Lists. Multiple estimation of significance values can increase type I errors (i.e., rejecting the null hypothesis \(H_0\) when \(H_0\) is true). Thus, in tables 3–6 we also present \(p\)-values corrected using a Benjamini–Hochberg procedure (BH; Benjamini and Hochberg, 1995) to control for potentially false discovery rates (FDR), the expected proportion of ‘discoveries’ (rejected null hypothesis \(H_0\)) that might be false (incorrect rejection). Nevertheless, this type of correction incurs reduction in power. Using this kind of procedure for the more detailed studies would have implied a lower probability of finding significant results, increasing the risk of type II errors, sometimes to an unacceptable level (not rejecting \(H_0\) when \(H_0\) is false) (Nakagawa, 2004). Ecological results suggested by BH \(p\)-values were generally similar to those indicated by the \(p\)-value of Fisher’s tests. Thus, in the text we only detail and discuss the results from Fisher’s tests.

Results

Traceability

Although all 66 species from our sample were included in the global and national Red Lists, only 17–63 of these species were included in the European and sub-national lists (table 1). The classification criteria were clearly presented in almost all studied lists, the exceptions being the lists for Midi–Pyrénées, Provence–Alpes–Côte–d’Azur and Bretagne. Sources and clear links between data in assessments and sources were simultaneously available for global and European Red Lists but not for national and sub-national lists. Nevertheless, sources (but not links) were given for the Bourgogne, Limousin, Île–de–France, Midi–Pyrénées and Provence–Alpes–Côte–d’Azur sub-national lists (table 2). Accessibility and language constrained the literature review. The analysis of the types of information sources was based on 76.1% of sources (303 identifiable sources out of 398) for the global Red List, 91.2% of sources (1078/1182) for the European Red List and 100% of sources for sub-national lists (3–97 sources, depending on the
The links between information and sources were analyzed on the basis of (1) the traceable and accessible sources that determined the conservation status for the global Red List and (2) the relevant sources for France that were traceable and accessible in the European Red List. In the current global Red List, 48.9% (65/133) of citations were untraceable or not reported in the bibliography section. Of the 68 traceable sources (51.1% of sources) we managed to obtain 49 (72.1%), thereby allowing us to analyze 51% (177/347) of all links between information and source used in the evaluation of the conservation status of the species from our sample. For the European Red List, the sources cited for France represent 11.2% (132) of the whole bibliography for the study species. In this sample, 3.8% (5/132) of citations were untraceable. Of the 127 sources of identifiable literature (96.2% of sources), 111 were accessible and obtained (87.4%), thereby allowing us to analyze 83.2% (559/672) of the links between information and source cited for France.

These inequalities in information availability determined which comparisons between lists and scales could be performed.

**Conservation status**

The proportion of conservation statuses attributed varied between the Red Lists at different spatial scales (fig. 1, table 3). Overall, Red Lists at larger scales were associated with more ‘least concern’ statuses and fewer ‘threatened’ statuses. Nevertheless, European and national lists did not differ significantly and proportions of statuses between lists at the same spatial scale did not show any significant differences when examined directly.

Nonetheless, the comparisons of the two global lists, before and after the update at the end of 2015, with the sub–global lists did not give identical results. Overall, the pre–update global Red List exhibited fewer threatened statuses than the European and national lists; however, these differences disappeared when we compared them with the post–update global Red List. In addition, the pre–update global Red List exhibited few more differences than the post–update global Red List in comparison with the sub–national Red Lists.

**Red List criteria**

The proportion of the Red List criteria that was applied varied between the lists at different spatial scales (fig. 2, table 4). Criterion ‘A’ (reduction in population size) was used significantly more in the assessment of species at larger scales (global and European lists) than in national and sub–national lists. On the other hand, criterion ‘D’ (small regional population) was used more on sub–national lists than in national, European and global Red Lists.
Table 3. Results of the comparison of the proportions of different IUCN statuses on Red Lists of birds. N, number of species; p, p–values of Fisher’s test; BHp, p–values corrected using the Benjamini–Hochberg procedure; OR, odds ratio; BL, European Red List from Birdlife International, SN.SN, all comparisons of sub–national Red lists. (For other abbreviations see table 1).

|                | LC status | NT status | VU status | EN status | CR status |
|----------------|-----------|-----------|-----------|-----------|-----------|
|                | N         | p         | BHp OR    | p         | BHp OR    | p         | BHp OR    | p         | BHp OR    |
| IUCNo–IUCNd    | 66        | –         | –         | –         | –         | –         | –         | –         | –         |
| IUCNo–BL.Eur   | 63        | 0.004     | 3.284     | –         | –         | 0.009     | 0.018     | 0.087     | –         |
| IUCNo–BL.EU27  | 63        | 0.001     | 5.326     | –         | –         | 0.000     | 0.000     | 0.053     | –         |
| IUCNd–BL.Eur   | 63        | –         | –         | –         | –         | –         | –         | –         | –         |
| IUCNd–BL.EU27  | 63        | –         | –         | –         | –         | 0.011     | 0.015     | 0.219     | –         |
| IUCNo–UICN     | 66        | 0.002     | 6.023     | –         | –         | 0.030     | 0.006     | 0.000     | –         |
| IUCNd–UICN     | 66        | –         | –         | –         | –         | –         | –         | –         | –         |
| IUCNo–Al       | 34        | 0.000     | 23.557    | –         | –         | –         | –         | 0.024     | 0.000     |
| IUCNo–Bo       | 28        | 0.000     | 25.502    | –         | –         | 0.023     | –         | 0.000     | –         |
| IUCNo–Br       | 41        | 0.000     | 20.588    | –         | –         | –         | –         | 0.025     | 0.000     |
| IUCNo–Ce       | 32        | 0.000     | 23.557    | –         | –         | –         | –         | 0.025     | 0.000     |
| IUCNo–IDF      | 28        | 0.000     | Inf       | –         | –         | –         | –         | –         | –         |
| IUCNo–PL       | 33        | 0.002     | 9.739     | –         | –         | –         | –         | –         | –         |
| IUCNo–PACA     | 34        | 0.000     | 28.693    | –         | –         | 0.002     | 0.040     | 0.000     | –         |
| IUCNo–Li       | 31        | 0.000     | 31.453    | –         | –         | –         | –         | 0.021     | 0.000     |
| IUCNo–MIP      | 17        | 0.004     | 15.311    | –         | –         | –         | –         | 0.021     | 0.000     |
| IUCNo–LaR      | 32        | 0.000     | 13.512    | –         | –         | 0.020     | –         | 0.000     | –         |
| IUCNd–Al       | 34        | 0.011     | 5.401     | –         | –         | –         | –         | 0.025     | 0.000     |
| IUCNd–Bo       | 28        | 0.009     | 5.799     | –         | –         | –         | –         | 0.025     | 0.000     |
| IUCNd–Br       | 41        | –         | –         | –         | –         | 0.025     | –         | Inf       | –         |
| IUCNd–Ce       | 32        | 0.011     | 5.401     | –         | –         | –         | –         | 0.025     | 0.000     |
| IUCNd–IDF      | 28        | 0.027     | 5.434     | –         | –         | –         | –         | 0.025     | 0.000     |
| IUCNd–PL       | 33        | –         | –         | –         | –         | –         | –         | 0.025     | 0.000     |
| IUCNd–PACA     | 34        | 0.003     | 6.616     | –         | –         | 0.013     | –         | 0.086     | –         |
| IUCNd–Li       | 31        | 0.005     | 5.581     | –         | –         | –         | –         | 0.021     | 0.000     |
| IUCNd–MIP      | 17        | 0.004     | 15.311    | –         | –         | –         | –         | 0.021     | 0.000     |
| IUCNd–LaR      | 32        | 0.026     | 4.060     | –         | –         | 0.020     | –         | 0.000     | –         |
| BL.Eur–BL.EU27 | 63        | –         | –         | –         | –         | –         | –         | 0.024     | 0.000     |
| BL.Eur–UICN    | 63        | –         | –         | –         | –         | –         | –         | 0.024     | 0.000     |
| BL.Eur–Al      | 34        | 0.026     | 4.179     | –         | –         | –         | –         | 0.024     | 0.000     |
| BL.Eur–Bo      | 28        | 0.023     | 4.469     | –         | –         | –         | –         | 0.024     | 0.000     |
| BL.Eur–Br      | 41        | –         | –         | –         | –         | –         | –         | 0.025     | 0.000     |
| BL.Eur–Ce      | 32        | 0.026     | 4.179     | –         | –         | –         | –         | 0.025     | 0.000     |
| BL.Eur–IDF     | 28        | 0.027     | 5.434     | –         | –         | –         | –         | 0.025     | 0.000     |
Some significant differences appeared between Red Lists in adducing criteria B and C, no clear trends emerged when examining spatial scale. Overall, the proportions of adduced Red List criteria did not vary between lists at comparable spatial scales (global and European), although a few exceptions did occur between sub–national lists (table 4).

### Bibliographical categories

The Red Lists at different scales used different literature. For instance, Red Lists at the global scale were more based on scientific literature \textit{sensu stricto} (about 50\%) than European (about 10\%) and subnational Red Lists (0–6\%), which were, in turn, more based on grey literature (fig. 3, table 5). The Red List for geographical Europe was more based on scientific literature (12\%) and less on grey literature (53\%) than the list for the EU27 (7\% and 60\%, respectively). The European lists were more based on expert opinion (16–18\%) and less on books (15–17\%) than the global lists (0.4–1.6\% and 24.5–26.1\%, respectively).

At comparable spatial scales, the updated global Red List was based on a smaller proportion of scientific literature (47\%) than the previous version (56\%) (fig. 3). There were also a few other differences between sub–national Red Lists related to their use of books and grey literature as sources (table 5).

### Citation categories

The global and European Red Lists (the only ones in which the links between information and source could be analysed) had different proportions of citation categories (fig. 4, table 6). The results revealed that ‘clearly supported’ assertions were significantly more common in the global Red List (83\%) than in the
European list (32%). ‘Ambiguous’ and ‘not supported’ assertions were both significantly more abundant in the European Red List (both 34%) than in the global Red List (13 and 4%, respectively) (fig. 4; table 6). Comparisons of citations in Red Lists according to the bibliographic category of the original sources highlighted the fact that ‘clearly supported’ assertions were more abundant in citations of books (86.5%) and grey literature (79.1%) from the global Red List while ‘no supported’ assertions were more abundant in citations of books (30.4%) and grey literature (31.1%) of the European Red List.

In the global Red List, there were no significant differences in citation categories between the citations from different bibliographical category. However, in regard to the European List, ‘ambiguous’ assertions were more frequently linked to grey literature than to books and ‘not supported’ assertions were more frequently linked to books than to scientific articles and grey literature.

Discussion

We conducted this study for game bird species in France and therefore the applicability of the results to other species or other geographical areas still remains an open question. Despite the huge volume of work that was required for this study, the sample size was still on occasions a limiting factor when attempting to unravel some of the less evident differences, for instance in comparisons at equivalent geographic scales. Notwithstanding this limitation, this study highlights clear trends in scale–dependent patterns.

IUCN standards and transparency

We found clear differences in the transparency and the traceability of the assessment processes used for Red Lists at different geographic scales. Although all the Lists considered in this study were certified by logos and labels that are directly linked to the IUCN guidelines (or indirectly through the IUCN French committee), national and sub–national lists in France were the product of data and processes that could not be verified and were not presented in a transparent and accessible way. Thus, the national and sub–national Red Lists in France do not fully comply with the standardized processes ‘to be applied without deviation or modification, if regional Red List authorities wish to state that their assessment follows the IUCN system’ (IUCN, 2012). Thus, we need additional information to be able to fully understand
Table 4. Results of comparison between Red Lists for the criteria leading to threatened status of birds: N, number of criteria; p, p-values of Fisher’s test; BHp, p-values corrected using the Benjamini–Hochberg procedure; OR, Odds ratio; A, population reduction; B, geographic range; C, small population size and decline; D, very small or restricted regional population; BL, European Red List from Birdlife International. (For other abbreviations see table 1).

| Criteria A        | Criteria B        | Criteria C        | Criteria D        |
|-------------------|-------------------|-------------------|-------------------|
| N, number of criteria | p, p-values of Fisher’s test | BHp, p-values corrected using the Benjamini–Hochberg procedure | OR, Odds ratio; A, population reduction; B, geographic range; C, small population size and decline; D, very small or restricted regional population; BL, European Red List from Birdlife International. (For other abbreviations see table 1). |
| IUCNo–IUCNd 3–5 | – | – | – | – |
| IUCNo–BL.Eur 3–10 | – | – | – | – |
| IUCNo–EU27 3–18 | – | – | – | – |
| IUCNd–EU27 5–10 | – | – | – | – |
| IUCNo–UICN 3–13 | 0.000 0.002 Inf | – | – | – |
| IUCNd–BL.Eur 5–10 | – | – | – | – |
| IUCNd–BL.EU27 5–18 | – | – | – | – |
| IUCNo–IDF 3–13 | 0.003 0.011 Inf | – | – | – |
| IUCNo–Ce 3–10 | 0.001 0.005 Inf | – | – | – |
| IUCNo–Al 3–11 | 0.004 0.011 Inf | – | – | – |
| IUCNo–LaR 3–11 | 0.001 0.007 Inf | – | – | – |
| IUCNo–Li 3–13 | 0.000 0.000 Inf | – | – | – |
| IUCNd–Bo 5–13 | – | – | – | – |
| IUCNd–PL 5–8 | 0.036 0.042 Inf | – | – | – |
| IUCNd–IDF 5–7 | 0.028 0.036 Inf | – | – | – |
| IUCNd–Ce 5–10 | 0.015 0.023 Inf | – | – | – |
| IUCNd–Al 5–11 | 0.038 0.041 Inf | – | – | – |
| IUCNd–LaR 5–11 | 0.015 0.026 Inf | – | – | – |
| IUCNd–Li 5–13 | 0.010 0.020 Inf | – | – | – |
| BL.Eur–BL.EU27 10–18 | – | – | – | – |
| BL.Eur–UICN 10–13 | 0.000 0.000 Inf | – | – | – |
| BL.EU27–UICN 18–13 | 0.000 0.000 Inf | 0.037 | 0.000 0.010 0.020 0.000 0.037 | 0.000 |
| BL.Eur–Bo 10–13 | 0.000 0.000 Inf | – | – | – |
| BL.Eur–PL 10–8 | 0.000 0.000 Inf | – | – | – |
| BL.Eur–IDF 10–7 | 0.000 0.000 Inf | 0.015 | 0.000 0.000 |
| BL.Eur–Ce 10–10 | 0.000 0.000 Inf | – | – | – |
| BL.Eur–Al 10–11 | 0.000 0.000 Inf | – | – | – |
| BL.Eur–LaR 10–11 | 0.000 0.000 Inf | – | – | – |
| BL.Eur–Li 10–13 | 0.000 0.000 Inf | – | – | – |
| BL.EU27–Bo 18–13 | 0.000 0.000 Inf | – | – | – |
| BL.EU27–PL 18–8 | 0.000 0.000 Inf | – | – | – |
| BL.EU27–IDF 18–7 | 0.000 0.000 Inf | 0.003 0.042 0.000 | – | 0.015 0.016 0.000 |
| BL.EU27–Ce 18–10 | 0.000 0.000 Inf | – | – | – | 0.000 0.000 0.000 |

Tabla 4. Resultados de las comparaciones entre Listas Rojas de los criterios que determinan que las aves están amenazadas: N, cantidad de criterios; p, valor p de la prueba de Fisher; BHp, valor p corregido según el procedimiento de Benjamini–Hochberg; OR, razón de momios; A, reducción de población; B, distribución geográfica; C, tamaño de población pequeño y en decrecimiento; D, población local muy pequeña o limitada; BL, Lista Roja europea de Birdlife International. (Para las otras abreviaturas véase tabla 1).
how lists including national and sub–national lists in France that deviate from the IUCN standards on transparency benefit from the labels that link them, either directly or indirectly via national committees, to the IUCN system. Likewise, on–line searchable databases including all the details of the data and assessments accessible via the Internet (e.g. http://www.iucnredlist.org, available for the global Red List) should be required for all sub–global Lists. This improvement in transparency should be a priority for national and sub–national lists that are, at least in France, the less transparent ones.

Conservation status

Our results for game birds in France agreed with the predictions derived from previous studies (Gärdenfors et al., 2001; Keller et al., 2005) and highlight the fact that Red Lists at smaller geographical scales frequently give higher threatened statuses than those at larger scales. The reason for this might be local variability in the status of species when compared to conservation status at larger scales. Species may exhibit a threatened status first at a local level prior to exhibiting threatened status at a global level or even, in occasion,
several species may be less threatened at local level than globally (Szabo et al., 2012). Nevertheless, the averagely higher threatened statuses of birds on Red Lists at smaller scales might also be linked to the risk of pessimistic assessments at regional level owing to over–narrow scale–dependent geographical focus (Keller et al., 2005; IUCN, 2012). As seen above, a risk of subjectivity in the regional adjustment process has already been identified (Keller and Bollmann, 2004; Rodríguez et al., 2004; Keller et al., 2005), in addition to risks of ‘scale–’ and ‘edge–effect’ (Keller et al., 2005; Milner–Gulland et al., 2006; Brito et al., 2010). Thus, these potential risks and the observed not supported data that was included in Red Lists does not rule out the possibility that the higher threatened statuses of Red Lists at smaller scales in France might be due, at least in part, to methodological factors or potential bias. Thus, methodological improvements aimed at reducing the risk of subjectivity in the second step of the IUCN regional guidelines and at avoiding the ‘scale–’ and ‘edge–effect’ in assessments are needed to strengthen the robustness of the Red Lists.

Red List criteria

Our results also match the predictions derived from previous articles on regional IUCN Red List assessments (Eaton et al., 2005; Keller et al., 2005), and underscore the fact that assessments of threatened status in lists at smaller scales in France were predominantly based on the criterion ‘D’ (small regional population), while criteria linked to reductions of population (criterion ‘A’) were predominant in Red Lists at larger (European and global) scales. This scale–dependent characteristic of the assessment process may highlight and emphasize ‘scale–’ or ‘edge–effects’ in regional Red Lists, as has previously been reported (Eaton et al., 2005, Keller et al., 2005). Variation in the most commonly adduced criteria may reflect what data are available to assess species at the scale in question, and so data availability may hamper the feasibility and reliability of assessments at spatial scales that are too small.

Bibliographical categories

Our results concur with the predictions for bibliographical categories in Red Lists at different geographic scales and reveal that Red Lists at smaller scales were in general based on grey literature, while global Red Lists were more based on scientific articles. These results underlined the greater risk of subjectivity in Red Lists at smaller geographical scales. This risk is even greater in the European Red...
Table 5. Results of comparisons between Red Lists of birds for bibliographical categories: N, number of references; p, p-values of Fisher's test; BHp, p-values corrected using the Benjamini–Hochberg procedure; OR, odds ratio; A, scientific literature sensu stricto; B, books and referenced academic publications; C, grey literature; BL, European Red List from Birdlife International. (For other abbreviations see table 1).

| N       | A          | B           | C           | D           |
|---------|------------|-------------|-------------|-------------|
|         | p BHp OR   | p BHp OR    | p BHp OR    | p BHp OR    |
| IUCNo–IUCNd | 237–303   | 0.037 0.037 | 1.449       | – – – – – – |
| IUCNo–BL.Eur| 237–1,078 | 0.000 0.000 | 8.689       | 0.009 0.009 | 1.584 0.000 | 0.000 0.198 | 0.000 0.000 | 0.000 0.022 |
| IUCNo–BL.EU27| 237–615   | 0.000 0.000 | 16.514      | 0.003 0.004 | 1.795 0.000 | 0.000 0.158 | 0.000 0.000 | 0.000 0.020 |
| IUCNd–BL.Eur| 303–1,078 | 0.000 0.000 | 5.997       | 0.001 0.002 | 1.724 0.000 | 0.000 0.287 | 0.000 0.000 | 0.000 0.088 |
| IUCNd–BL.EU27| 303–615   | 0.000 0.000 | 11.410      | 0.000 0.000 | 1.953 0.000 | 0.000 0.229 | 0.000 0.000 | 0.000 0.078 |
| IUCNo–Bo   | 237–97    | 0.000 0.000 | 19.247      | 0.000 0.000 | 4.152 0.000 | 0.000 0.040 | – – – – – – |
| IUCNo–PACA| 237–6     | 0.008 0.016 | Inf         | – – – – – – | 0.002 0.004 | 0.048       | – – – – – – |
| IUCNo–IDF | 237–23    | 0.000 0.000 | Inf         | – – – – – – | 0.000 0.000 | 0.084       | – – – – – – |
| IUCNo–Li  | 237–5     | 0.018 0.030 | Inf         | – – – – – – | 0.006 0.010 | 0.060       | – – – – – – |
| IUCNd–Bo   | 303–97    | – – – – – – | 0.000 0.000 | 13.310 0.000 | 0.000 4.521 | 0.000 0.000 | 0.000 0.058 |
| IUCNd–PACA| 303–6     | 0.033 0.047 | Inf         | – – – – – – | 0.006 0.009 | 0.069       | – – – – – – |
| IUCNd–IDF | 303–23    | 0.000 0.000 | Inf         | – – – – – – | 0.000 0.000 | 0.121       | – – – – – – |
| IUCNd–MiP | 303–3     | – – – – – – | 0.019 0.048 | 0.000       | – – – – – – | – – – – – – |
| IUCNd–Li  | 303–5     | – – – – – – | – – – – – – | 0.018 0.023 | 0.086       | – – – – – – |
| BL.Eur–BL.EU27| 1,078–615 | 0.000 0.000 | 1.904       | – – – – – – | 0.028 0.028 | 0.797       | – – – – – – |
| BL.Eur–Bo | 1,078–97  | – – – – – – | 0.009 0.030 | 2.627 0.000 | 0.000 0.200 | 0.000 0.000 | 0.000 18.207 |
| BL.Eur–PACA| 1,078–6   | – – – – – – | – – – – – – | – – – – – – | – – – – – – | – – – – – – |
| BL.Eur–IDF| 1,078–23  | – – – – – – | – – – – – – | – – – – – – | – – – – – – | – – – – – – |
| BL.Eur–MiP| 1,078–3   | – – – – – – | 0.005 0.025 | 0.000       | – – – – – – | – – – – – – |
| BL.Eur–Li | 1,078–5   | – – – – – – | – – – – – – | – – – – – – | – – – – – – | – – – – – – |
| BL.EU27–Bo| 615–97    | – – – – – – | 0.040       | – – – – – – | 2.318 0.000 | 0.000 0.252 | 0.000 0.000 0.000 20.642 |
| BL.EU27–PACA| 615–6    | – – – – – – | – – – – – – | – – – – – – | – – – – – – | – – – – – – |
| BL.EU27–IDF| 615–23   | – – – – – – | – – – – – – | – – – – – – | – – – – – – | – – – – – – |
| BL.EU27–MiP| 615–3    | – – – – – – | 0.004 0.040 | 0.000       | – – – – – – | – – – – – – |
| BL.EU27–Li | 1,078–5   | – – – – – – | – – – – – – | – – – – – – | – – – – – – | – – – – – – |
| Bo–PACA  | 97–6      | – – – – – – | – – – – – – | – – – – – – | – – – – – – | – – – – – – |
| Bo–IDF   | 97–23     | – – – – – – | – – – – – – | – – – – – – | – – – – – – | – – – – – – |
| Bo–MiP   | 97–3      | – – – – – – | 0.001 0.010 | 0.000 0.004 | 0.040 | Inf         | – – – – – – |
| Bo–Li    | 97–5      | – – – – – – | – – – – – – | – – – – – – | – – – – – – | – – – – – – |
| PACA–IDF | 6–23      | – – – – – – | – – – – – – | – – – – – – | – – – – – – | – – – – – – |
| PACA–MiP | 6–3       | – – – – – – | 0.048       | – 0.000 0.048 | Inf         | – – – – – – |
| PACA–Li  | 6–5       | – – – – – – | – – – – – – | – – – – – – | – – – – – – | – – – – – – |
| IDF–MiP  | 23–3      | – – – – – – | 0.022       | – 0.000 0.032 | Inf         | – – – – – – |
| IDF–Li   | 23–5      | – – – – – – | – – – – – – | – – – – – – | – – – – – – | – – – – – – |
| MiP–Li   | 3–5       | – – – – – – | – – – – – – | – – – – – – | – – – – – – | – – – – – – |
List than in global Red Lists due to its higher reliance on expert opinions, even though the revisions of the IUCN assessment criteria in recent decades have been explicitly oriented towards reducing subjectivity (Mace and Lande, 1991; Rodrigues et al., 2006). Thus, these differences may weaken the reliability of regional Red Lists compared to global Red Lists. Furthermore, the greater dependence on scientific literature in the previous global Red List than in the current global Red List highlight the need for greater attention to be paid (1) to ensuring that the IUCN system, designed to provide comprehensive, scientifically rigorous information, is reliably applied at global and regional levels (IUCN, 2012), and (2) to preventing the current uncertain assessment processes at regional level, which use predominantly grey literature and expert opinions, from being increasingly applied at a global scale. If society collectively wants science–based and reliable Red Lists at small spatial scales and if peer–reviewed literature is (as it currently is) the widely accepted strategy for ensuring quality control in scientific research (Ferreira et al., 2015), the publication of small–scale studies in peer–reviewed journals will be necessary even despite the inherent difficulties of the publication process. The further integration of these potential needs by the editors of scientific journals might help promote more reliable Red Lists at regional scales in the future.

Citation categories

Finally, our results agreed with the predictions on the citation categories and show that 'not supported' assertions were frequently linked to grey literature and books. Nevertheless, our results also highlighted the high degree to which the type of Red List affects these results. The global Red List is predominantly based on assertions 'clearly supported' by the cited references. 'Ambiguous' and 'not supported' citations were a minority (less than 20 %) in the global Red List but were a majority (more than 65 %) in our sample from the European Red List. Numerous citations of books in the European Red List were particularly questionable, for instance, old references that were cited (e.g. from 1964, 1977 or 1994) as support for assertions regarding recent short–term population trends (e.g. for Alauda arvensis, Aythya ferina, Galinago gallinago, Limosa limosa, Numenius arquata, Tetrao urogallus). These results underline the fact that numerous assertions that were either ambiguous or not supported by the cited books and grey literature were included in regional assessments. Thus, the studied Red Lists, focused at different geographic scales, exhibit significant heterogeneity in both fundaments and reliability. This highlights the need for additional reviewing processes for sub–global assessments and, in particular, checks of
the accuracy of links between citations and primary sources. Furthermore, the relative lack of data and the difficulties of the publication process may potentially increase the temptation to use grey literature in Red List assessments. Nevertheless, our results reveal that such practices increase the risk of inclusion of ambiguous and not supported data in regional assessments. Compensating for a lack of data (Butchart and Bird, 2010) by using grey literature may misrepresent the situation of species that would otherwise be regarded as ‘data deficient’ according to IUCN guidelines (DD conservation status), and thus may reduce the visibility of the ignored information that could justify financial support for additional research and species monitoring. Consequently, conservation decisions may be associated with an ‘assessment dilemma’: should we report data deficiencies strictly to highlight the need for further research and thus have to confront potential delays in evidence–based management?, or should we use all available information to promote reactive management, albeit at the risk of using not–supported and/or unreliable data in assessments, thereby jeopardizing the credibility of Red Lists and reducing the visibility of needs to improve species monitoring? Further studies directly focused on this dilemma could have constructive implications for the monitoring and consensual conservation of species.

**Conclusion**

This study mainly revealed information about monitoring schemes, data collection and availability at different spatial scales in birds from Europe and France in particular. Sub–national to global Red Lists differed in regard to (i) the reported conservation status, (ii) the transparency and traceability of assessments, (iii) the most commonly adduced criteria, (iv) the categories of the sources synthesized during assessments, (v) and the reliability of assertions compared to cited references. Such variability between lists in terms of both data and transparency confirms that the sources used in the global Red List were cited as reliably as usual in ecological sciences (Todd et al., 2007, 2010). However, there were many ambiguous

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**Table 6. Results of comparisons of citation categories between the Global and European Red Lists of birds: Bc, bibliographical category; Ns, number of sources; p, p-values of Fisher’s test; BHp, p–values corrected using the Benjamini–Hochberg procedure; OR, odds ratio; IUCN; Global Red List from the International Union for Conservation of Nature; BL, European Red List from Birdlife International; A, scientific literature sensu stricto; B, books and references academic publications; C, grey literature.**

|                | Clear support |            | Ambiguous |            | No support |            |
|----------------|---------------|------------|-----------|------------|------------|------------|
|                | p             | BHp        | OR        | p          | BHp        | OR         |
| IUCN–BL        |               |            |           |            |            |            |
| All            | 0.000         | 0.000      | 10.629    | 0.000      | 0.000      | 0.296      |
| A oct–26       | –             | –          | –         | –          | –          | –          |
| B 89–164       | 0.000         | 0.000      | 14.483    | 0.002      | 0.003      | 0.314      |
| C 67–357       | 0.000         | 0.000      | 8.343     | 0.001      | 0.002      | 0.328      |
| IUCN           |               |            |           |            |            |            |
| A.B oct–89     | –             | –          | –         | –          | –          | –          |
| B.C 89–67      | –             | –          | –         | –          | –          | –          |
| A.C oct–67     | –             | –          | –         | –          | –          | –          |
| BL             |               |            |           |            |            |            |
| A.B 26–174     | –             | –          | –         | –          | –          | 0.009      |
| B.C 174–357    | –             | –          | –         | 0.011      | 0.033      | 0.599      |
| A.C 26–357     | –             | –          | –         | –          | –          | 1.656      |
and not supported citations on the European Red List and unverifiable assessments on the national and sub-national lists. These results thus open the door for further analysis and improvements of the reliability of Red Lists at regional levels and for other taxa to strengthen evidence-based wildlife management and avoid the decrease in the credibility and prestige of IUCN-based Red Lists (Mrosovsky, 1997; Mrosovsky and Godfrey, 2008) in the eyes of researchers, the general public and other stakeholders.

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