Deriving “potential target values” of PCDD/F in animal feed: the role of livestock at the interface between feed and food chain

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Abstract
Linking derivation of potential target values of PCDD/Fs in animal feed with risk assessment for consumer protection is a challenge when tolerable weekly intake (TWI) and transfer factors from feed to food are considered. Generally, maximum values for feed and food are set separately without considering the feed and the food producing animal as an important factor along the food chain from farm to fork. Levels of contaminants in feed can accumulate in animals and their products effect consumers at the end of the food chain. Hence, the process of setting legal maximum levels of contaminants should account for transfer from feed consumed by food producing animals into animal products for human consumption. Here, we calculated potential target values of PCDD/F in feed to ensure that animal products such as milk from dairy cows, eggs from laying hens and pork and pork products from fattening pigs are safe for human consumption. In our approach, we calculated potential target values of PCDD/Fs in animal feed using transfer factors for PCDD/F-TEQs from feed to milk fat, eggs fat, and fat in pork and pork products, taking into account the tolerable weekly intake derived by European Food Safety Authority. We assumed equal proportions of WHO-PCDD/F-TEQ and WHO-PCB-TEQ in feed. Potential target values of PCDD/F in feed are expressed as the quantity of toxicologically evaluated PCDD/Fs, expressed in WHO toxic equivalents (WHO2005-PCDD/F-TEQ) per kg feed with 12% moisture. In the current approach, derived values would be 10–54 times lower than the current legal maximum level of 0.75 ng WHO2005-PCDD/F-TEQ per kg feed (12% moisture), according to Directive 2002/32/EC as amended.

Keywords Livestock · Feed and food chain · Dioxins · WHO2005-PCDD/F-TEQ · Transfer factor · Human risk assessment

1 Introduction

Polychlorinated dibenzodioxins and -furans (PCDD/Fs), commonly termed “dioxins”, and dioxin-like polychlorinated biphenyls (DL-PCBs) are two- or three-ring structures that can be chlorinated to varying degrees. DL-PCBs can have up to ten chlorine atoms substituting hydrogen atoms, whereas PCDD/Fs can have up to eight (WHO 2010). PCDD/Fs refer to two groups of tricyclic planar compounds (PCDDs and PCDFs) for which, depending on the number of chlorine atoms and their positions at the rings, a total of up to 75 PCDDs and 135 PCDFs so called ‘congeners’ can occur. The 17 compounds, which are chlorinated in the 2, 3, 7, 8 position, share the same mode of action, are highly toxic and at the same time persistent. Thus, they accumulate in the adipose tissue of animals and humans and are therefore considered relevant for human (and animal) health. For risk assessment purposes, toxicity equivalency (TEQ) was developed to describe the cumulative toxicity of complex mixtures of these compounds. The most toxic congener, 2, 3, 7, 8-TCDD, is assigned a value of 1. The toxicity equivalency factors (TEFs) for the individual congeners are between 0 and 1, indicating the magnitude of their toxicity in relation to 2, 3, 7, 8-TCDD. The TEF values including DL-PCB were first proposed by the WHO in 1997 (WHO1998-TEQs; van den Berg et al. 1998) and were re-evaluated by a WHO working group in 2005 (WHO2005-TEQs; van den Berg et al. 2006, in the following referred as “WHO2005-TEQ”).

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The amount of a congener in a sample is multiplied by the respective TEF. The sum of all resulting products is a measure of the dioxin-related toxicity of the congeners in the sample, the so-called WHO-TEQs.

PCDD/Fs and PCBs are very stable against chemical and microbiological degradation and therefore persistent in the environment from where they are transferred into the food chain. Hence, food is considered to be the major source of human exposure to PCDD/Fs and PCBs, with the exception of specific cases of accidental or occupational exposure (Travis and Hattemer-Frey 1991; Fries 1995; Windal et al. 2010; EFSA 2012a, 2018). In contrast to PCBs, PCDD/Fs have never been produced on an industrial scale and have no technological use. They are formed unintentionally in a number of industrial and thermal processes as unwanted and often unavoidable impurities or reaction by-products (EFSA 2012a, 2018).

Because of their lipophilic properties, these substances preferentially accumulate in adipose tissues. Thus, foods of animal origin showed highest contribution to human exposure to PCDD/Fs and PCBs. In 2001, the Scientific Committee on Food (SCF) which was the predecessor of the European Food Safety Authority (EFSA), derived a tolerable weekly intake (TWI) of 14 pg/kg bw as health based guidance value (HBGV) for PCDD/Fs and DL-PCBs. In 2018, the Panel on Contaminants in the Food Chain (CONTAM) of EFSA, re-assessed the health risk posed by exposure to PCDD/Fs and DL-PCBs. The new EFSA-TWI of 2 pg/kg bw is seven times lower than the TWI of 2001. It is based essentially on a study conducted on young men, in whom exposure of PCDD/Fs and DL-PCBs on sperm quality and sperm count was examined (EFSA 2018). The CONTAM panel estimated the weekly intake of dioxins and DL-PCBs for various age groups based on the consumption and occurrence data in Europe. The panel concludes that exposure in European consumers exceed the new TWI considerably in all age groups (EFSA 2018).

Both, maximum levels in feed and food for PCDD/Fs and DL-PCBs are generally not set based on toxic effects. Derivation of maximum levels of contaminants in feed (and food) is often performed without consideration HBGVs. Consumption of food derived from food producing animals may lead to exposure exceeding the TWI (e.g. for dioxins/DL-PCB) although feed was legislative compliant with European maximum levels according to Directive 2002/32/EC (2002). Moreover, such feed fed to livestock can cause food of animal origin that exceeds the legal maximum levels for food (according to Commission Regulation (EC) No. 1881/2006 as amended by Commission Regulation (EU) No 1259/2011 (e.g. Lorenzi et al. 2020; Spitaler et al. 2005). Consequently, substance transfer along the feed-to-food-chain and consideration of HBGV should be part of a practical approach when deriving maximum levels of undesirable substances in feed.

By doing so, the challenge is to find maximum levels which are both practically applicable hence avoiding feed waste, and protective for animals and consumers at the end of the food chain. (Very) low values might not be of practical use if either background levels are already at the same level or even higher and/or for analytical reasons e.g. if they could not be reproducibly determined. This might be the case when they are below or in the range of detection or quantification limits.

In this article, we demonstrate an approach that considers several steps in the food chain, regarding PCDD/Fs in feed which is consumed by food producing animals, transferred to food for human consumption. The calculation of such values—in our paper referred to as “potential target values” was exemplarily performed for PCDD/F WHO2005-TEQ in feed for dairy cows, fattening pigs, and laying hens. The aim was that consumption of milk from dairy cows, eggs from laying hens and pork and pork products from fattening pigs will not exceed EFSA’s TWI for PCDD/F-DL-PCB. The quandary of the attempt should be part of a broader discussion.

1.1 Objective

The objective of this work is the derivation of potential target values of PCDD/Fs in feed for dairy cows, fattening pigs, and laying hens, using the new tolerable weekly intake (TWI) derived by EFSA (2018). The values are exemplarily derived for PCDD/Fs only (all congeners). Potential target values for DL-PCBs or WHO2005-PCDD/F-DL-PCB-TEQ were calculated separately for transfer into milk and results were compared. There is less literature reporting transfer factors for DL-PCBs at steady state conditions (Amutova et al. 2021; Lorenzi et al. 2020; VDI 2019). PCDD/F concentrations in the total feed ration should consider potential target values for farm animals that serve to protect these animals thereby ensuring that foods derived from them are safe for human consumption (VDI 2005, 2019). The derived potential target values in feed are indicated as the quantity of toxicologically evaluated PCDD/Fs, expressed in WHO toxic equivalents (WHO2005-PCDD/F-TEQ) per kg feed (88% dry matter and relative to a feed with 12% moisture).

2 Material and methods

2.1 Calculation of transfer factors

The calculation of transfer factors (TF) is presented in Table S1 Supplementary Material. Briefly, Lorenzi et al. (2020) reported concentration of PCDD/F-TEQ, DL-PCB-TEQ and of PCDD/F + DL-PCB-TEQ after addition of defined concentrations of PCDD/F und PCB congeners in feed. The reported transfer into milk (pg/g fat) at the end of
the exposure period (day 49, steady state conditions) was divided by the concentration in feed (pg/g feed). Uptake from feed and excretion into milk is very variable between congeners. We accounted for this variation by using TEQ values for the sum of PCDD/F congeners. Likewise, TFs were calculated from the feeding study performed by Hoogenboom et al. (2015) using PCDD/F-TEQ and DL-PCB-TEQ values in maize silage and reported concentration in milk at day 29 of exposure (Table S2, Supplementary Material). We used the mean TF of both studies for PTV calculation as described below.

To calculate PTV transfer from feed for laying hens into eggs, the study performed by Stephens et al. (1995) was identified as most appropriate. TFs for excretion of PCDD/F congeners were calculated similarly using concentrations given in feed (including soil, pg/g) and eggs (pg/g fat), applying WHO2005 TEF values (Table S3, Supplementary Material).

Finally, we calculated TFs for transfer from feed for fattening pigs into pork on information obtained from the study performed by Spitaler et al. (2005). Transfer into belly fat was calculated for PCDD/F-TEQ based on two study groups receiving 0.75 and 2 pg/g feed, respectively. Lower concentrations in feed resulted in higher TFs (1.1 and 2.2, respectively). The higher value rounded to 2.0 was applied for PTV calculation (Table S4, Supplementary Material).

2.2 Assumptions

In order to link derivation of potential target values of PCDD/Fs in animal feed with risk assessment for consumer protection, the following assumptions were taken into account:

- Equal proportions of WHO-PCDD/F-TEQ and WHO-PCB-TEQ in feed were assumed. As the TWI established by EFSA (2018) is of 2 pg WHO2005-PCDD/F-PCB-TEQ/kg bw/week, a “tolerable daily intake” (TDI1) of WHO-TEQ is assumed to be 0.14 pg/kg body mass for PCDD/F (50% of TWI on a daily base).
- Kinetics of the individual congeners are different resulting in different transfer rates from feed into food (e.g. congeners 2, 3, 7, 8-TCDD, 1, 2, 3, 7, 8-PCDD and 2, 3, 4, 7, 8-PCDF account for as much as 71% of the total WHO-TEQ in cows’ milk, 67% in pork and 62% in eggs as calculated from German Food Monitoring data 2007–2011). We calculated potential target values in feed (PTVF) based on WHO2005-PCDD/F TEQ.
- The TFs, defined as quotient of concentration in pg/g fat in animal product under steady state conditions and concentration in pg/g feed; dry matter, of the three WHO2005-PCDD/F-TEQ were:
  - 3.3 to 5.6 from feed to milk fat (Hoogenboom et al. 2015; Lorenzi et al. 2020). A mean TF of 4.5 were taken forward for calculation of PTVF.
  - 1.9 from feed to eggs fat was taken for the calculation (Stephens et al. 1995, calculated as WHO2005-PCDD/F-TEQ).
  - 1.1 to 2.2 from feed to fat in pork and pork products, depending on concentration of PCDD/F-TEQ in feed (Spitaler et al. 2005). The mean value was taken forward for calculation (calculated as WHO2005-PCDD/F-TEQ).

The most recent and appropriate studies were selected if they met the following criteria: a comprehensive description of the experimental setup, duration of exposure which showed steady state conditions, studies considering all 17 congeners and, with species in questions with good productivity. We only used studies that allows us to calculate TFs based on PCDD/F-TEQ and—in the case of transfer into milk—additionally for DL-PCB-TEQ. The TF of WHO2005-DL-PCB-TEQ was 5.3 to 7.5 from feed to milk fat (Hoogenboom et al. 2015; Lorenzi et al. 2020). The mean transfer factor for the calculation of PTVF for DL-PCB was 6.4. It is generally accepted that food is the major source of exposure to PCDD/F for consumers, where milk fat accounts for 30%, egg fat for 6% and fat from pork and pork products up to 5% of the total amount of WHO-TEQ main food group levels (Schwarzet al. 2014). A “daily fat consumption” per person per day of 40 g in milk of dairy cow, 3.8 g in eggs and 11.3 g from pork and pork products (Table 1) was used for calculation (Schwarz et al. 2012; Statista 2020). Based on all parameters described and considered (Table 1), potential target values of PCDD/Fs in feed were calculated in order to comply with the TWI derived by EFSA (2018).

3 Results

Concepts describing transfer from feed into animal products varies in literature. In our calculation, we selected studies that gave sufficient information on study design and

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1 A TDI for WHO2005-PCDD/F-PCB-TEQ does not exist. The “TDI” value used in the calculation was taken from the TWI divided by 7 to obtain a reference value on the daily basis. Fifty percent of the values were used, assuming equal contribution of exposure of DL-PCB and PCDD/F.

2 Huwe and Smith (2005) reported comparable bioconcentration factors for transfer from contaminated mineral supplement in feed into milk from dairy cows of 5.5 and 5.9 at steady state.
outcome in order to calculate TFs. Lorenzi et al. (2020) and Hoogenboom et al. (2015) calculated carry-over rates from the quotient of concentration in milk fat (pg/g fat) * fat yield (g/d) and the concentration in feed (pg/kg feed) * feed intake (g/d). We use the term “transfer factor” as quotient from the concentration of PCDD/F-TEQ and DL-PCB-TEQ in milk fat and the concentration in feed at steady state conditions. Hence, fat yield and feed consumption as given in the respective studies are covered but are no direct parameters in our calculation of $PTV_F$. The same applies for the calculation of TF for animal feed into egg’s fat and pork fat, respectively (Stephens et al. 1995; Spitaler et al. 2005). The term “transfer factor” is used throughout the manuscript and calculated to account for accumulation in fat over time. In the literature, some authors used “bio-concentration factor” (BCF) with the same definition (e.g. Lorber et al. 2000; Huwe and Smith 2005) while others defined it differently (e.g. Amutova et al. 2021).

### 3.1 General equation for the calculation of potential target values ($PTV_F$) in feed

Based on the described assumptions, tolerable levels $PTV_F$ of PCDD/F in animal feed were calculated using the following equations:

$$PTV_F = \frac{TDI \times \text{Human bw} \times \text{Percentage of PCDD/F intake with food}}{\text{Daily Fat Consumption} \times \text{Transfer Factor}}$$

(TDI): pg/kg bw/d (EFSA 2018), [bw]: kg (EFSA 2012b). [Percentage of PCDD/F intake with food]: without unit (decimal digit; Schwarz et al. 2014). [Daily fat consumption]:kg/d (Schwarz et al. 2012; Statista 2020). [Transfer factor]: without unit (Quotient pg/g fat in diet/ pg/g feed at steady state).

Equation 1: $PTV_F$ for feed for dairy cows in pg/kg dry matter:

$$PTV_F = \frac{0.14 \times 70 \times 0.3}{0.04 \times 4.5}$$

Equation 2: $PTV_F$ for feed for laying hens in pg/kg dry matter:

$$PTV_F = \frac{0.14 \times 70 \times 0.06}{0.0038 \times 1.9}$$

Equation 3: $PTV_F$ for feed for fattening pigs.

$$PTV_F = \frac{0.14 \times 70 \times 0.05}{0.0113 \times 2}$$

Therefore, based on the assumptions above, the potential target values for feed for dairy cows, laying hens and fattening pigs are (Table 2):

| Feed materials of plant origin | Calculated potential target values in ng WHO-PCDD/F-TEQ/kg relative to a feed with a moisture content of 12% | Legal limit: maximum content in ng WHO-PCDD/F-TEQ/kg relative to a feed with a moisture content of 12% |
|---|---|---|
| Dairy cows | 0.014 | 0.75 |
| Laying hens | 0.072 | |
| Fattening pigs | 0.019 | |
PTVF\textsubscript{Dairy cow} = 0.016 ng WHO-TEQ/kg dry matter (equivalent to 14 pg/kg feed with 12% moisture).

PTVF\textsubscript{laying hens} = 0.081 ng WHO-TEQ/kg dry matter (equivalent to 72 pg/kg feed with 12% moisture).

PTVF\textsubscript{Fattening pigs} = 0.022 ng WHO-TEQ/kg dry matter (equivalent to 19 pg/kg feed with 12% moisture).

Taking into account the TWI of 2 pg WHO-PCDD/F-TEQ/kg bw for human exposure (by using the calculated “tolerable daily intake” (“TDI”) of 0.14 pg), the derived PTV\textsubscript{F} would be 10 to 54 times lower than the current legal maximum level of 0.75 ng WHO-TEQ per kg feed (12% moisture), according to Directive 2002/32/EC (2002).

In order to compare PTV\textsubscript{F} DL-PCB-TEQ with PTV\textsubscript{F} PCDD/F-TEQ, Eq. 1 was slightly adopted. We used TFs 7.47 (calculated from Lorenzi et al. 2020) and 5.32 (calculated from Hoogenboom et al. 2015) and the mean of both values (TF: 6.4; Table S1, S2 Supplementary Material) were included in our calculation (Eq. 1a). Percentage of DL-PCB exposure with dairy products is comparable to PCDD/F-TEQ and DL-PCB-TEQ (30% vs. 27%, resp.) (Schwarz et al. 2014). The same accounts for contribution of egg fat to the overall exposure to DL-PCB with food. A clearly higher contribution to overall food exposure to DL-PCB was only reported for pork (Schwarz et al. 2014). However, as no eligible studies were identified meeting our defined inclusion criteria, we did not calculate a PTV\textsubscript{F} values for feed for laying hens and for feed for fattening pigs based on DL-PCB-TEQ.

Equation 1a: PTV\textsubscript{F} for feed for dairy cows in pg/kg dry matter based on dl-PCB-TEQ:

$$PTV\textsubscript{F} = \frac{0.14 \ast 70 \ast 0.3}{0.04 \ast 6.4}$$

PTVF\textsubscript{F} = 11 pg/kg dry matter (equivalent to 10 pg/kg feed with 12% moisture).

The potential target value for feed for dairy cows based on DL-PCB-TEQ would be slightly lower compared to the PTV\textsubscript{F} based on PCDD/F-TEQ.

### 4 Discussion

In our approach, we calculated potential target values of PCDD/F in animal feed. Derived values should ensure safe levels in livestock products for human consumption. Here, we demonstrate the challenge to derive potential target values for PCDD/F in feed for dairy cows, fattening pigs, and laying hens, considering the transfer from animal feed to livestock products for human consumption under the provision to not exceed the new TWI for PCDD/F (EFSA 2018). With the calculated low potential target values of the TWI (EFSA 2018), it was difficult to comply with the PCDD/F levels in the feed ration, because the calculated levels were 10–54 times lower as the current legal value (0.75 ng/kg feed with 12% moisture).

#### 4.1 Analytical challenge

According to Commission Regulation (EU) 2017/771, the limit of quantification (LOQ) for a confirmatory method should be approximately 1/5 of the legal maximum level (Dir 2002/32/EC), to ensure the reliability of the performance results and the whole test method at low concentrations (EURL 2016). However, the potential target values calculated here by using the lowered TWI of 2 pg/kg bw and week, are 2.1–10.7 times below the currently required LOQ for PCDD/Fs in feed raw materials and compound feed. Hence, lowering legal maximum levels in feed would also require improvement of analytical performance. Commission Regulation (EU) No 277/2012 already mentions the dependence of maximum levels on analytical performance. It states that setting lower maximum levels was not possible regarding the sensitivity of currently available methods of analysis at that time.

#### 4.2 Compliance with legal maximum levels in feed does not mean compliance with legal maximum levels in food

Commission Regulation (EU) No 277/2012 further addresses the conflict regarding food of animal origin which exceeds the applicable maximum levels set by Commission Regulation (EC) No 1881/2006, even though maximum levels in feed comply with Directive 2002/32/EC. Lorenzi et al. (2020) selected the contamination level in order to comply with legal maximum levels for feed. The TEQ ratio between DL-PCB and PCDD/F was 3 to 1 in order to resemble the average TEQ ratio found in local forages. In their study, milk of exposed cows exceeded the maximum limit set for the sum of PCDD/Fs and DL-PCBs (5.5 pg TEQ/g fat) after one week of exposure whereas WHO\textsubscript{2005}-PCDD/F-TEQ alone never exceeded legal maximum levels but reached established action levels of 1.75 pg TEQ/g fat) at day 21 of exposure. Hoogenboom et al. (2006) described exceedance of maximum levels in eggs in chickens fed with feed containing TEQ levels at or just above (0.76 ng/kg feed) the current EU maximum levels of 0.75 ng TEQ/ kg feed. Our approach focusses on health-based guidance values instead of legal maximum levels in food. It demonstrates that potential target values are lower than current legal maximum levels for feed. In addition, maximum levels for food and feed are generally set based on occurrence data for feed and food and/or on analytical practicality. Thus, a comparable approach could also be used to calculate PTV in food. Such approach must include dietary exposure assessment which would be an interesting point for further studies.
4.3 Variability of transfer factors

Kinetics of the individual congeners are different resulting in various transfer rates from feed into food. Calculation based on total PCDD/F-TEQ in feed and the respective food (fat based) is possible if concentrations in food and feed are reported based on the sum of PCDD/F-TEQ. Regarding the literature values of TF or BCF on single congeners, our TFs are well within the range of reported values (e.g. Huwe and Smith 2005 and references cited therein). The most conservative approach would have been to use a TF of the toxicologically most relevant congeners with the highest transfer of toxicity. A recent meta-analysis shows mean transfer rates of 34.0 ± 6.3% and 39.1 ± 12.6% for TCDD from feed into milk and eggs, respectively (Amutova et al. 2021). For DL-PCB-126, the toxicologically most relevant DL-PCB congener, the transfer rates were 40 ± 11.4% and 37.7 ± 19.6% from feed into milk and eggs, respectively. In our calculation, we used TFs which accounted for accumulation of PCDD/Fs and DL-PCBs in animal products. If transfer rates are used, the equations must be adopted.

We calculated TFs and exemplarily considered only one livestock product, milk, egg, and pork, respectively. In dairy cows we only looked at the transfer from feed into milk, whereas transfer into meat was disregarded. As the liver also accumulates PCDD/F, there is an uncertainty regarding congener specific accumulation in fat, liver or kidney (e.g. Huwe and Smith 2005) and with that also the consumer’s safety regarding liver or kidney consumption. However, accumulation in liver of cows is less important compared to e.g. in sheep (EFSA 2011; Fernandes et al. 2011). Furthermore, most important for human exposure is the consumption of milk and meat products. Hence, this uncertainty may lead to a slight overestimation but is justifiable regarding the theoretical and simplified approach of our calculation. A refinement could be achieved by modelling e.g. mass balances.

Comparable or slightly higher values would be achieved if PTVF are calculated based on the sum of PCDD/F and DL-PCB-TEQs.

Various definitions and values are reported for the transfer from animal feed to livestock products for human consumption. Based on TF or BCF values summarized by Huwe and Smith (2005) for single PCDD/F congeners, values appear to depend on several factors, e.g. feed material, concentrations in feed, study duration etc. The range for the toxicologically most relevant congeners (2, 3, 7, 8-TCDD and 1, 2, 3, 7, 8-PeCDD) has been reported to be between 1.12 and 8.99 (mineral supplement), with most values being between 4 and 8. The values are comparable to the TF we used in our calculation. For example, an increase of TFs (e.g. twofold) would result in a PTVF half as high as the reported ones and vice versa.

4.4 Steady state conditions in transfer studies

Concentrations in animal products were chosen if steady state conditions were reported (Lorenzi et al. 2020; Hoogenboom et al. 2015). Those were 49 and 29 days of exposure for transfer into milk, respectively. The transfer from feed into eggs took 178 days of exposure (Stephens et al. 1995), and a 12-week exposure period (study duration 18 weeks) for the transfer from feed into pork of fattening pigs (Spitaler et al. 2005). Normally, feed intake at steady state applies to adult animals with good productivity, except for fattening pigs whose daily feed consumption steadily increases from app. 1 kg to 3 kg in accordance to its live weight gain of up to 800 g per day until reaching 100–120 kg during the fattening period (Jeroch et al. 2008). Hence, achievement of stable conditions is difficult to achieve in fattening pigs (Amutova et al. 2021). If exposure periods are not sufficiently long to achieve steady state conditions, accumulation of PCDD/F would be underestimated. For instance, Hoogenboom et al. (2015) reported “nearly steady state conditions” whereas Lorenzi et al. (2020) reported steady state conditions. We used the mean of TFs in our calculation which may lead to a slightly underestimation if steady state conditions in the study by Hoogenboom et al. was not achieved (TF were lower than in study by Lorenzi et al. 2020). Feed intake is not expected to have a big influence at steady state. The daily fat production and animals body weight were not directly calculated, but indirectly through the calculation of fat concentration in feed and milk, egg, and pork meat, respectively. Calculation of TF at steady state conditions in the fat of the regarded livestock product implies that both feed consumption and fat production in animals have no impact on PCDD/F concentration in fat.

4.5 Variability in PCDD/F and DL-PCB congeners transfer

There is a discrepancy between the percentages of overall PCDD/F contribution of congeners in food and feed compared to the contribution of congeners to the overall PCDD/F toxicity. Data shows that 1, 2, 3, 4, 6, 7, 8, 9-OCDD; 1, 2, 3, 4, 7, 8, 9-HpCDD; and OCDF were quantitatively the
dominant congeners in feed and food of animal origin. However, they were the least relevant to toxicity, due to their low WHO-TEF value and their low transfer rates (e.g. Fries et al. 1999; Malisch 2000; Brambilla et al. 2008; EFSA 2012a; EFSA 2018). Recently, Lorenzi et al. (2020) confirmed in a controlled feeding study that OCDD characterized the milk profile in untreated animals, or in treated animals before and during the clearance phase. Around 30% of PCDD/F in milk derived from unexposed cows, compared to approx. 4% at steady state in milk derived from exposed cows. OCDD, HpCDD, and 2, 3, 4, 7, 8-PeCDF accounted for around 52% of total PCDD/F in milk. In contrast, 2, 3, 7, 8-TCDF contributed to max. 0.5% of PCDD/F concentration in milk in unexposed cows but was responsible for around 10% of overall toxicity. In exposed cows, contribution increased from approx. 0–2.4% (day 49, end of exposure phase). However, calculation of TF from PCDD/F-TEQ in food and feed considers the different contribution of congeners to PCDD/F toxicity.

4.6 Variable exposure of PCDD/F feed on livestock

Generally, it could be assumed that livestock is not long time exposed to feed with homogenously levels of dioxins. Thus, realistic exposure would include high and low exposure scenarios. On the other hand, we showed that derived PTV$_F$ are low and in the magnitude of background levels for feed. For example, median levels of PCDD/F-TEQ (17 congeners) in 206 samples of feed (roughage, fresh forage—fresh grass etc.—compound feed) were 0.03 ng/kg (88% DM) and clearly below maximum level of 0.75 ng/kg. Our derived PTV$_F$ of 0.016—0.081 ng PCDD/F WHO$_{2005}$-TEQ/kg DM are comparable to mean concentrations of 0.03 – 0.09 ng WHO$_{2005}$-TEQ/kg DM (median of all feed samples: 0.02 ng/kg WHO-TEQ/kg 88% DM) in which none of the samples exceeded legal ML (Federal Ministry of Food and Agriculture 2009). Mueller et al. (2021) found very low concentrations of PCDD/F-TEQ single congeners in feed from grassland (closely to or below LOQ: LOQ: not reported). Mean concentration of PCDD/F- WHO$_{2005}$-TEQ in grassland feed (unwashed) was 0.09 ng/kg DM (range: 0.053 to 0.191 ng/kg DM). Concentrations of single congeners were very low, often at or below LOQ (LOQs were not reported in the publication). Current background levels for roughage and compound feed in German monitoring data were between 0.01 and 0.008 ng WHO-PCDD/F-TEQ/kg 88% DM, respectively (CVUA Freiburg 2019). As described for feed, MLs for food items are generally set based on occurrence data (e.g. P95) rather than on data on toxicity. If livestock is exposed to PCDD/F in feed compliant with derived PTV$_F$, exceedence of maximum levels in food are not expected. Transfer studies showed that PCDD/F levels in milk, eggs and pork meat may exceed MLs or action levels if content in feeds is below or at the current ML for feed (roughage, compound feed) of 0.75 ng TEQ/kg (Hoogenboom et al. 2015; Lorenzi et al. 2020; Spitaler et al. 2005).

4.7 Uncertainty in the derived TWI for PCDD/F + DL-PCB-TEQ

Finally, there are also uncertainties concerning the TWI of 2 pg/kg bw/week due to the contribution of exposure of congeners to PCDD/F and DL-PCB TEQ, especially of PCB 126. The latter has the highest toxicity potency (WHO$_{2005}$-TEF: 0.1) regarding DL-PCBs. TEF factors will be re-assessed by an WHO expert group following a peer review of the relative effect potencies of PCDDs, PCDFs and PCBs (EFSA 2021$^4$). If this assessment results in different TEFs it will influence overall assessment of dioxin exposure and toxicity. For instance, Lorenzi et al. (2020) demonstrated an overall contribution of PCB 126 to DL-PCB TEQ between 74 and 91% in milk of exposed and unexposed cows, respectively. PCB 126 is the DL-PCB contributing most to the current intake of PCDD/F and DL-PCB-TEQ with food (EFSA 2018). If TEF of PCB 126 would result in a lower value than the pattern of contribution to overall DL-PCB toxicity (and also PCDD/F + DL-PCB TEQ) may change significantly.

4.8 Other uncertainties

In our approach, an underestimation can occur if PCDD/F transfer from soil to livestock is disregarded for animals in free-range production, which applies in particular to poultry and grazing cattle (e.g. Schwind et al. 2010). Some authors included concentrations of soil and feed in their studies (e.g. Hoogenboom et al. 2015). In addition to feed, livestock in free-range husbandry can also take up PCDD/F via grazing or soil ingestion (Schoeters and Hoogenboom 2006). In particular, the intake of contaminated soil leads to a significant accumulation in animals and transfer to food of animal origin (Schulz et al. 2005). Any additional intake of PCDD/F through sources other than feed will lead to lower potential target values for feed, as calculated. Moreover, there are no current studies on PCDD/F and PCB levels in feed of different feed categories such as compound feed, roughage and forages. The last national survey of these contaminants in feed and food of animal origin dates back to 2009 in Germany (Federal Ministry of Food and Agriculture 2009). Therefore, we propose to perform a new feed survey, because current contamination of feed with PCDD/Fs and PCBs might have changed over time. In addition, analytical performance could also be improved since the most recent publication of maximum levels in feed.

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$^4$ https://www.efsa.europa.eu/en/call/peer-review-database-relative-effect-potencies-pcdds-pcdfs-and-pcbs-and-preparing-review-use, Accessed 20.10.2021.
For derivation of potential target values in feed, equal proportions of PCDD/Fs and PCBs in feed were used for calculation. However, a 3:1 ratio for DL-PCB and PCDD/Fs in feed and food is more realistic (e.g. Lorenzi et al. 2020; Schwarz et al. 2014; Hoogenboom et al. 2004; Huwe and Smith 2005). Nevertheless, there would be in the same magnitude if the calculation was performed including DL-PCB PTVF in dairy cow feed for PCDD/F + DL-PCB-WHO2005-TEQ.

We based our calculation of PTVF on adult consumers. However, children are likely to be more vulnerable to PCDD/F + DL-PCB exposure. The TWI by EFSA was based on serum concentration in 5-year old boys after a modelled external exposure in order to derive the TWI. For this reason, the TWI also covers exposure to children. If we use a body weight of 12 kg for children as a default value for children aged of 1–3 years (EFSA 2012b), in our calculations PTVF would be lower than in 70 kg adults (approx. by factor 6). Thus, our derived PTVF may underestimate the risk of PCDD/F exposure in children. Sources and directions of uncertainties of PTVF derivation are shown in Table 3.

### 5 Conclusions

In conclusion, the described approach could contribute to the discussion about the need for an integrative approach considering derivation of maximum levels in feed and food, and thereby also regarding risk assessment for human health. Livestock has an important role along the food chain from farm to fork. Levels of contaminants in feed can accumulate in animals and their products reach consumers at the end of the food chain. Several steps along the food chain should be regarded when setting maximum levels for feed because transfer from feed into animals and further into food for human consumption are relevant with regard to consumer’s safety, especially for lipophilic contaminants. This approach might serve as a management tool to better integrate the setting of maximum levels in feed and food with health based guidance values for human risk assessment. Our PTVF should be regarded as “target values” rather than proposals of MLs for feed. This could be compromised by aiming for avoidance of feed and food waste and by analytical practicality. However, levels in feed might have been decreased during the last ten years or so. Thus, a new feed survey would help to reflect the current situation regarding concentrations of PCDD/F and DL-PCB in feed.

### Supplementary Information

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### Authors contributions

MJAC: performed research, analyzed data, wrote the paper, JK, BW: designed the study, wrote the paper.

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