Method Article

A data independent acquisition all ion fragmentation mode tool for the suspect screening of natural toxins in surface water

M. Picardo, O Núñez, M. Farré *

Water and Soil Quality Research Group, IDAEA-CSIC, c/Jordi Girona 18-26, Barcelona 08034, Spain

ABSTRACT

Among natural freshwater pollutants, cyanotoxins, mycotoxins, and phytotoxins are the most important and less studied. Their identification in surface water is challenging especially cause of the lack of standards and established analytical parameters. Most target methods focus one or a single group of compounds with similar characteristics. Here we present an AIF fast method for the tentative identification of natural toxins in water. Respect to the previous method [1], it offers higher performances for the acquisition of unknown compounds at low levels for higher number of analytes.

The key aspects of the method are:
• The qualitative screening DIA-AIF workflow using Q Exactive Orbitrap. Both targeted and suspect screening bases have been combined with online databases and suspect list to retrieve candidates as suspect natural toxins and their metabolites or degradation products.
• The in-silico analysis of mass spectrums allowed a fast structural characterization.
• The workflow has been finally applied to real samples coming from the Czech Republic, Italy, and Spain allowing the determination of 17 suspect natural toxins, 4 of them confirmed. None toxin passed the limit of 1 μg/L taken from the legislation applied for microcystin LR and arbitrarily extended to all toxins.

© 2021 The Authors. Published by Elsevier B.V. This is an open access article under the CC BY-NC-ND license (http://creativecommons.org/licenses/by-nc-nd/4.0/)

ARTICLE INFO

Method name: Data Independent Acquisition All Ion Fragmentation mode
Keywords: AIF, DIA, HPLC-HRMS/MS, QExactive Orbitrap, Tentative identification
Article history: Available online 22 February 2021

* Corresponding author.  
E-mail address: mfuqam@cid.csic.es (M. Farré).
Specifications table

| Subject Area | Method name | Name and reference of original method | Resource availability |
|--------------|-------------|---------------------------------------|-----------------------|
| Chemistry    | Environmental Analytical Chemistry | Picardo M., Sanchís, J., Nuñez O., Farré M. Suspect screening of natural toxins in surface and drinking water by high performance liquid chromatography and high-resolution mass spectrometry. Chemosphere Volume 261, December 2020, 127888 | Compound Discoverer 3.1 (ThermoFisher); MzCloud; MetFrag; Xcalibur |

Method details

Common approaches for the analysis of natural toxins in surface waters rely on solid-phase extraction as a sample preparation protocol followed by target analysis with Data Dependent Acquisition methods for a limited number of compounds. Most methods are specifically designed for a group of toxins with similar parameters or a single compound, depending on its physicochemical characteristics. However, the prioritization of natural toxins and their degradation products in the surface water environment is of increasing importance due to their different eco-toxicological properties [2].

The need for identification protocols is critical, especially considering the low availability of certified standards. Among them, targeted approaches are generally used to analyze known chemicals of interest while non-targeted approaches are more challenging. This is due to the need for identification and structure characterization protocols that require the use of multiple instruments (NMR and IR) which usually are not available or highly expensive.

High-Resolution Mass Spectrometry (HRMS) based on high-resolution instruments such as Orbitrap, QTOF, and FTICR are helping to fulfill the need for reliable identification methods, providing sensitive fragmentation spectra (MS/MS) for the identification of known and unknown compounds [3]. HRMS provides a high amount of information for characterization and identification purposes (molecular formula, isotopic patterns, double bond equivalents) comparing the experimental results with online or in-house databases of chemical compounds. Tandem mass spectrometry and the consequent fragmentation spectra are mandatory to achieve a tentative structural characterization. In these regards, the data-acquisition methodology used to acquire MS/MS spectrums is of critical consideration that influences the type of data generated, and the choice of which method to use is largely dependent on the aim of the approach. Among them, Data-Dependent Acquisition (DDA), Single Reaction Monitoring (SRM), and Data-Independent Acquisition (DIA) are the most used (Fig. S1 of the Supporting Information).

This work aims to introduce the All Ions Fragmentation (AIF) acquisition approach as a suspect screening method for a wide range of natural toxins in surface water. The AIF acquisition for all theoretical fragment ions was used to acquire the entire MS/MS spectrum with no precursor preselection. Data processing and information extraction required the use of various bioinformatics tools to deconvolute complex mass spectra, using data from prior experiments in DDA mode to generate spectral libraries that were used in the interrogation of DIA data [4]. The objectives can be resumed in (i) develop a robust workflow for the determination of natural toxins in surface water samples using the AIF mode; (ii) provide a reliable workflow to describe how to process the acquired data, (iii) demonstrate the advantages to use this approach as a tentative identification protocol for the screening of natural toxins using real samples.

A Q-Exacte Orbitrap was used to obtain the full scan and MS/MS spectrums with the AIF mode. Data mining was then carried out using Compound Discoverer using a published suspect list with 2384 natural toxins [1] and the online databases Chemsider and MassBank [5] and also with fragmentation prediction tools such as MetFrag. The “Fish score” option was used to structurally characterize the MS/MS patterns. Finally, 24 natural toxins have been tentatively identified from surface water samples coming from three sampling sites in Europe.
Standard solutions

Table S1 of the supporting information reports the standards used for method optimization. Compounds 1–5, 7–12, 14, 15, and 22–26 were supplied from Merck (Darmstadt, Germany). Compound 6 was supplied from Santa Cruz Biotechnology (Dallas, TX, USA). 16–21 were from Cyano (Cyanobiotec GmbH, Berlin, Germany). Methanol (MeOH), acetone, and acetonitrile (ACN) HPLC grade were from Merck (Darmstadt, Germany). HPLC grade water was from Baker (Madrid, Spain). Fortified samples (5 mL each) with the 23 compounds (Table S1) at a concentration of 1 µg/L were prepared in both HPLC water and artificial freshwater (AFW) adding 10 % of MeOH, to reach the initial chromatographic conditions, and to simulate the presence of matrix interference. Samples were mixed with magnetic bar stirring at 200 rpm and letting set for an hour to ensure the good mixing of the standards. To prepare the AFW we followed the description of Lipschitz and Michel [6], the organic matter was simulated by adding 10 mg/L of humic acid of technical grade from Sigma-Aldrich (reference 53680). The method optimization was carried out analyzing the standard solution in pure HPLC and artificial water, previously mixed for an hour at 25 °C and processed as reported below.

Sample preparation

Sample preparation was previously reported by Picardo et al. [1]. Briefly, intracellular toxins were released by sonication for 20 min and further filtered with a glass fibre (GF/B) microfiber filter grade (Sigma Aldrich, Steinheim, Germany). Solid-phase extraction (SPE) consisted of a 3 mL cartridge filled with 200 mg of porous graphitized carbon (PGC) (Sigma Aldrich, Steinheim, Germany) and 200 mg of Polypropylene polymeric phase Bond-Elut PPL (PPL) (Agilent, Santa Clara, CA, USA) separated by a Teflon frit. The third sorbent was the Oasis HLB plus, 225 mg (Waters Corporations, Milford, MA) connected at the end of the cartridge. Conditioning required 10 mL of methanol followed by 10 mL of water. 100 mL of sample was loaded at a constant flow rate (2 mL/min) using a vacuum manifold. After the procedure, analytes were eluted in backflush with 15 mL of water/methanol 20:80 (v/v), 15 mL of methanol, and 15 mL of acetone/methanol 1:1 (v/v). Solvents were warmed at 45 °C before elution. The eluate was concentrated to approximately 100 µL using a teardrop ampoule connected to a vacuum evaporator (rotavapor) and re-dissolved to 1 mL of mobile phase (approx. 0.9 mL of ACN 10% acidified at 0.1 % of formic acid) in a tared vial. Finally, 20 µL of samples were injected in the HPLC-HRMS/MS instrument.

Liquid chromatography-mass spectrometry

Chromatographic separation followed the same parameters reported by Picardo et al. [1]. Briefly, the separation was performed with an Acquity UPLC System (Waters, Milford, MA, USA) using a Lichrosphere column, 125 mm × 2 mm i.d., 5 µm (Merck, Barcelona, ES). 20 µL of samples were injected, the constant flow rate was 250 µL/min. Mobile phases were water (A) and acetonitrile (B) acidified with 0.1 % of FA. The gradient was 0–3 min, 10% of B; at 3–13 min B was increased to 90% and kept at a constant concentration from 13 to 15 min; 15–16 min B decreased to 10%; 16–20 min equilibration at 10% B. Total run time was 20 min. The analysis was performed using a Q-Exactive™ Orbitrap mass spectrometer (Thermo Fisher Scientific, San Jose, CA, USA). Samples were acquired in Full Scan and AIF mode in positive (+) mode in a range from 75 to 1100 m/z. Collision Induced Dissociation (CID) was obtained using a normalized collision energy of 35 eV. The mass spectrometer parameters are reported in Table 1.

Suspect screening workflow

After the acquisition, spectral data were further processed to tentatively identify suspect natural toxins with the screening approach as reported below.

In silico processing

No inclusion list of precursors ions was used in the acquisition method however, the suspect list reported by Picardo et al. [1] was used in combination with the in silico tools for the data
processing. Blanks of the entire procedure were processed to exclude background noise. Then, raw files from Orbitrap were uploaded to Compound Discoverer 3.1 (Thermo Fischer Scientific, San Jose, CA, USA) and processed with the Environmental Untargeted Metabolomics workflow. Here, peak alignment, unknown compound detection, compound grouping across all samples, elemental composition prediction, and chemical background hiding (using blank samples) were applied with a mass error of < 5 ppm. Finally, a tentative list of various compounds is displayed. Table 2 reports the parameters used for the compound identification with Compound Discoverer 3.1

**Structural identification**

The structure elucidation of the compounds in the candidate list was based on accurate mass data processing using Compound Discoverer 3.1 nodes. The molecular ions, potassium, sodium, and ammonium adducts and their transitions were used as identification parameters. Fish scoring node was applied to obtain the predicted structures of the transitions of each precursor selected in the full scan spectrum under the same retention time. MzLogic node was then used to compare the experimental and theoretical fragments contained in MzCloud. The spectrums have been submitted to MzCloud online search to obtain the corresponding similarity score (SS). Compounds with SS lower than 70% were discarded. Furthermore, MetFrag [7] was the last step for structural identification. Here, the MS/MS spectrums and the relative intensities have been uploaded. Here candidates have been retrieved using the molecular formula, neutral mass with a mass error of 5 ppm using the KEGG database. Only the first 10 candidates with a similarity score higher than 0.9/1.0 have been considered as valid candidates for the last step. Finally, each suspect natural toxin that fulfilled the requirements, was checked with Xcalibur (Thermo Fischer Scientific, San Jose, CA, USA) to control the elution profile and the retention times overlapping of the precursors and their transitions.

**Confirmation**

Confirmation was not possible for all the compounds due to the lack of standards. However, for the ones available the same procedure was carried out to obtain the AIF spectra from the standards to finally confirm the suspect compounds. (Level 1). Here identification levels from 1 to 5 were assigned to the suspect compounds, following what was previously reported by Schymanski et al. [8]. The lowest level 5 corresponded to the accurate mass, while level 4 was achieved using the spectral information to assign a molecular formula. Level 3 resulted at the end of the first identification step when the primary tentative candidate was proposed when existing some evidence to recognize a possible structure. Finally, levels 2 and 1 were achieved using databases reporting diagnostic evidence to assign an exact structure and using the standard respectively.
Table 2
Compound Discoverer 3.1 parameters for the peak alignment and identification.

| Workflow node          | Advanced Parameters       | Parameter               | Value         |
|------------------------|---------------------------|-------------------------|---------------|
| Select spectra         | Spectrum filter           | Min precursor Mass      | 75 Da         |
|                        |                           | Max precursor Mass      | 1100 Da       |
| Scan event filter      | Polarity mode             | Positive                |               |
|                        | Min Collision Energy     | 0                       |               |
|                        | Max Collision Energy     | 70                      |               |
|                        | Scan type                | Any                     |               |
| Align retention times  | Peak filter              | S/N threshold           | 3             |
|                        |                           | Max shift               | 2 min         |
| Find expected compounds|                           | Mass tolerance          | 5 ppm         |
|                        |                           | Intensity tolerance [%] | 50            |
|                        |                           | Intensity threshold [%] | 0.1           |
|                        |                           | S/N threshold           | 3             |
| Detect compounds       |                           | Mass tolerance          | 5 ppm         |
|                        |                           | Intensity tolerance [%] | 40            |
|                        |                           | S/N threshold           | 3             |
|                        |                           | Min peak Intensity      | 100000        |
|                        |                           | Ions Checked            |               |
|                        |                           | [M+ACN+H]^+, [M+H]^+,  |               |
|                        |                           | [M+K]^+, [M+NH_4]^+,    |               |
|                        |                           | [M+Na]^+                |               |
| Group Compounds        |                           | Min Elements count      | C, H, O       |
|                        | Consolidation            | Mass tolerance          | 5 ppm         |
|                        | Fragment data selection  | RT tolerance            | 1 min         |
|                        | Preferred Precursor Ions | [M+ACN+H]^+, [M+H]^+,  |               |
|                        |                           | [M+K]^+, [M+Na]^+ [M+NH_4]^+ |           |
| Search ChemSpider      |                           | Databases               | MassBank, Toxin, Toxin-target database, |
|                        | Search settings          | Search Mode             | By Formula and mass |
|                        |                           | Mass tolerance          | 5 ppm         |
|                        |                           | Max results per compound| 20            |
|                        |                           | Max predicted compounds | 3             |
| Search MzCloud         | General settings         | Library                 | Natural toxins |
|                        | DIA Search               | Use DIA scans for search| True          |
|                        |                           | Max isolation width     | 500           |
|                        |                           | Match activation type   | False         |
|                        |                           | Match Activation energy | Any           |
|                        |                           | Activation energy tolerance| 100          |
|                        |                           | Apply intensity threshold| False        |
|                        |                           | Match factor Threshold  | 10            |
| Search Mass list       | General settings         | Mass List               | In house suspect list |
|                        |                           | Use retention times     | False         |
|                        |                           | RT tolerance            | -             |
|                        |                           | Mass Tolerance          | 5 ppm         |

Application on real samples

The procedure was then applied to real samples coming from different sites in Europe. Briefly, 2 samples were from Piave River 46°10'12.6"N 12°15'58.2"E (Belluno, Italy), 3 from Sykovec (Tri Studne, Czech Republic), Brno Dam (49°13'58.1"N 16°31'03.3"E, Czech Republic) and Jedovnice (49°20'04.2"N 16°45'58.7"E, Czech Republic), respectively. 1 from Cardener River (41°40'48.2"N 1°50'39.1"E Barcelona, Spain). 1 L of surface water sample was collected in each point. All samples were processed in triplicate. Sampling was carried out between July and August were the highest biological activity was expected in the cited areas.
Optimization and suspect screening using AFW standards solutions

Unlike by what is generally applied with the ddMS\(^2\), all the characteristic MS/MS transitions that can ensure a positive assignment were considered. Fig. S2 reports the mass spectrum of the umbelliferone standard [M+H]\(^+\) m/z 163.0394 in artificial surface water after processing with Compound Discoverer 3.1 that matched the HR [M+H]\(^+\) with the Chemspider and the in-house suspect list using the exact mass of the precursor ion 163.0195 m/z with a maximum error of 5 ppm algorithm. The picture reports the Full Scan and the AIF spectrum. As expected, the fragmentation spectra differ between the one obtained with DDA since also other transitions coming from interfering ions are displayed. As shown in Fig. 1, the sensitivity increased when the spectrum is acquired in AIF mode providing higher intensities for the same fragments under the same experimental conditions (CE energy, concentration, MS parameters). A similar result was also highlighted by Sentandreu et al., [9] who reported that the breakdown pattern of previously isolated compounds and AIF patterns cannot be comparable. An over-breakdown is generally observed when AIF is applied retrieving a higher transition intensity rather than the molecular ion at the same collision energy. The tentative list obtained after the first analysis of AFW samples resulted in a list where appeared both compounds of interest and interferents. The noise was further hidden from the background using blanks. The first tentative structure was obtained using the “Flash scoring” employed to elucidate the structure of each transition in the MS/MS spectrum to predict in silico fragments based on the structure of the parent compound using a list of expected fragments reported in online databases. Then, the “mzLogic” algorithm compares the fragmentation patterns and the structures with MzCloud. Here, umbelliferone had a full match (100%) from our suspect list and a partial match in Chemspider with a score of 86%.

The most abundant fragment at 107.0492 m/z produced by the loss of –CO and –COH was observed followed in intensity by the 95.0492 and the 79.0180 m/z. However, during this experiment at least 5 positive fragments (green highlight in Fig. S3) were necessary to consider the compound as a tentative candidate. As a result of this processing, it is possible to observe that even if there were 80 unmatched transitions produced by interference, 16 were recognized as structural fragments. Spectra comparison depends on the Collision Energy applied. Here at 35 eV, umbelliferone structure C\(_9\)H\(_8\)O\(_3\) (7-hydroxycoumarine) was confirmed with a match score of 87.7% which is over the threshold required to accept a candidate to be further investigated. The MS/MS spectrum was also investigated using MetFrag [7] to predict the fragmentation and assign a formula for each transition. Here 2 candidates were displayed (umbelliferone and 4-hydroxycoumarin). The mass spectrum obtained a final similarity score of 1.0/1.0 for umbelliferone and 0.962/1.0 for the 4-hydroxycoumarin. A total of 29 fragments have been identified for the first compound and 27 for the second. Table 3 reports the structures, the formula, and the exact masses of the fragments considered for the tentative identification to level 2 of umbelliferone.

The final step to reach identification level 2 as reported by Schymansky et al. [8] was the manual check with XCalibur (Thermo Fischer Scientific, San Jose, CA, USA) to ensure the peak fitting under the same retention time of the precursor ion. Fig. 2 reports the MS\(^2\) spectra with the fragments considered for the tentative identification of umbelliferone. The intensities are higher with respect to the noise originated by the interferences allowing a clear recognition of the peaks. The procedure resulted in the overlap of retention times with the same peak shape and intensities of the MS/MS spectrums confirming the good performances obtained in the identification of spectral patterns. Here three fragments were considered as qualitative ions, with a mass error under 5 ppm, briefly: [C\(_6\)H\(_5\)O\(_2\)]\(^+\) 134.0368 m/z, 3.8 ppm; [C\(_7\)H\(_7\)O]\(^+\) 107.0495 m/z, 3.8 ppm; [C\(_6\)H\(_5\)O]\(^+\) 95.0495 m/z, 4 ppm; [C\(_7\)H\(_7\)O]\(^+\) 91.0546 m/z, 4.2 ppm; [C\(_6\)H\(_5\)O]\(^+\) 79.0546 m/z, 4.3 ppm. These steps were necessary to achieve the tentative identification level 2, however, the confirmation was only possible using standards.

**Confirmation**

The last step to confirm the suspect natural toxins to level 1, required the comparison with the standard. Fig. 3 shows the confirmation of the umbelliferone to the identification level 1. The standard solution at 1 μg/L in HPLC water was injected using the same acquisition method. As expected,
Fig. 1. Sensitivity comparison between SIM-ddMS/MS and AIF under the same experimental conditions.
Table 3
Fragmentation patterns recognised by MetFrag.

| Precursor | CompoundName  | Molecular formula | Identifier |
|-----------|---------------|-------------------|------------|
| Umbelliferone | C9H6O3 | C9H6O3 |

**Fragments**

| Formula | Mass | Peak m/z |
|---------|------|----------|
| \([\text{C5H2O-H}]^+\) | 77.00219 | 77.00256 |
| \([\text{C6H3+H}]^+\) | 77.0386 | 77.03893 |
| \([\text{C6H3+2H}]^+\) | 78.04643 | 78.04679 |
| \([\text{C5H3O}]^+\) | 79.01785 | 79.01823 |
| \([\text{C6H4+2H}]^+\) | 79.05426 | 79.0546 |
| \([\text{C5H4O}]^+\) | 81.03351 | 81.03388 |
| \([\text{C5H4O+2H}]^+\) | 83.04917 | 83.04953 |

(continued on next page)
| Precursor | Formula               | Mass     | Peak m/z  |
|-----------|-----------------------|----------|-----------|
|           | [C7H4]+H⁺              | 89.0386  | 89.03898  |
|           | [C7H4+H]+H⁺            | 90.04643 | 90.0468   |
|           | [C7H4+2H]+H⁺           | 91.05426 | 91.05464  |
|           | [C6H4O]+H⁺             | 93.03351 | 93.03397  |
|           | [C6H4O+H]+H⁺           | 94.04134 | 94.04169  |
|           | [C6H5O]+H⁺             | 95.04917 | 95.04956  |
|           | [C5H4O2+H]+H⁺          | 98.03625 | 98.03667  |
|           | [C8H5]⁻                 | 101.0386 | 101.03919 |
|           | [C7H5O]⁺                | 105.03351| 105.03397 |

(continued on next page)
| Precursor | Formula | Mass   | Peak m/z   |
|----------|---------|--------|------------|
|          | [C7H5O−H]+H+ | 107.04917 | 107.04959 |
|          | [C6H4O2]+ | 108.02059 | 108.02084 |
|          | [C6H4O2+2H]+H+ | 111.04408 | 111.04446 |
|          | [C8H6O−H]+ | 117.03351 | 117.03391 |
|          | [C8H6O]+ | 118.04134 | 118.04173 |
|          | [C8H6O]+H+ | 119.04917 | 119.04971 |
|          | [C9H5O−H]+H+ | 131.04917 | 131.04965 |
|          | [C8H6O2−H] + | 133.02842 | 133.0289 |
|          | [C8H6O2] + | 134.03625 | 134.0368 |

(continued on next page)
the typical fragment ions reported above and used as qualifier ions at m/z 134.0368, 107.0495, 95.0495, and 79.0546 were at the same retention time but with higher intensity. The signal was more intense due to the absence of interferents in the solution. The separation performance was comparable with the standards dissolved in AFW. The measured results were within the required limits for the identification of natural toxins in surface water samples. The same procedure was repeated with all the standards available. In Table 4 the results in AFW and HPLC water are reported. For each compound, more than 4 qualitative ions have been encountered in both AFW and HPLC water solutions. AFW samples presented as expected a lower signal suppressed by the most intense signals of the humic acids. However, the procedure allowed us to identify the standards and to validate the procedure for their determination. Quantitative validation was not included in this work since it is out of the aims.

**Surface water samples analysis**

Water samples coming from Italy, Spain, and Czech Republic were processed as described, performing the screening and the further identification and confirmation of different natural toxins. The pH was adjusted to 7.5 with formic acid 1.0 M, if necessary. For the one in which standards were not available identification levels (ILs) system was applied [8]. This ILs method has been used by other authors to identify low molecular mass molecules when using data-independent acquisition [1,10]. 138 compounds have been proposed as suspect candidates in the first identification step. However, only 27 were reported as suspect natural toxins, 3 (cotinine, abscisic acid, and ptaquilosin B) were false positives and 4 (methoxycoumarin, MC-LR, abietic acid, and umbelliferone) were confirmed comparing by standards (Table 5). For the compounds that had previous literature with mass spectra under similar conditions, the MS/MS interpretation was less time-consuming. For instance, the mass spectrum of azelaic acid matched with the one reported in MassBank [5]. Comparing the common fragments m/z = 83.08897, 97.10339, 103.05256, and 125.09818 were found in both spectra and the tentative identification level 2 was assigned. Then, the presence of suspect ptaquilosin B was also investigated. Ptaquiloside, a carcinogenic bracken fern toxin, is converted to the aglycone ptaquilosin B (PTB) in aqueous solutions due to the liberation of D-glucose to be then converted to pterosin B [11]. Here PTB was detected in the first identification step. However, the conversion rate of PTB depends on the temperature and the pH > 9. Here, since samples were frozen to -24 °C and the initial pH was 7.8 further investigation was required. Since D-glucose is released when converting ptaquiloside, its
Fig. 2. Manual check with Xcalibur for peak shape and retention time fitting of precursor and transitions.
Fig. 3. Mass spectrums of the standard umbelliferone and its fragmentation products.
| Compound         | Rt  | Precursor | Qi (1) | Structure  | Qi (2) | Structure  | Qi (3) | Structure  | Qi (4) | Structure  | Qi (5) | Structure  |
|------------------|-----|-----------|--------|------------|--------|------------|--------|------------|--------|------------|--------|------------|
| Ethoxycoumarin   | 11.2| 191.0698  | 107.0492 | C₇H₅O    | 95.0492 | C₆H₅O     | 163.039 | C₅H₄O₃    | 119.0492 | C₆H₅O     | 91.0543 | C₇H₇      |
| Methoxycoumarin  | 10.1| 177.0542  | 77.0386 | C₆H₅    | 162.0310 | C₆H₅O₃   | 106.0473 | C₇H₆O     | 121.0647 | C₆H₅O     | 134.0361 | C₆H₆O₂    |
| Abietic acid     | 1.5 | 303.3232  | 257.2269 | C₁₉H₂₀    | 121.1014 | C₆H₃       | 147.1171 | C₁₅H₁₀      | 287.2010 | C₁₉H₁₂O₂     | 241.1954 | C₈H₅O     |
| Aflatoxin B₁     | 9.8 | 313.0696  | 285.0763 | C₁₆H₁₉O₃  | 269.0449 | C₁₉H₁₀O₅  | 241.0499 | C₁₆H₈O₄     | 214.0627 | C₁₉H₁₀O₃     | 201.0913 | C₁₃H₆O₃     |
| Amygdalin        | 6.4 | 480.1483 [M+Na]+ | 85.0285 | C₆H₅O₂    | 107.0492 | C₆H₅O     | 325.11325 | C₁₆H₁₀O₁₀   | 163.0602 | C₆H₈O₅     | 127.0391 | C₆H₄O₃      |
| Anatoxin-a       | 1.6 | 166.1228  | 149.0964 | C₁₀H₁₃O   | 95.0493  | C₆H₅O     | 105.0700 | C₆H₈       | 91.0544  | C₇H₇       | 79.0544  | C₅H₇       |
| Atropine         | 6.8 | 290.1747  | 124.1120 | C₆H₄N     | 93.06989 | C₆H₅       | 103.0542 | C₄H₄       | 260.1644 | C₁₆H₂₂NO₂    | 142.1226 | C₅H₆NO      |
| B-Asarone        | 12.2| 209.1166  | 179.0705 | C₁₀H₁₉O₃  | 151.0756 | C₁₆H₁₀O₂   | 121.0649 | C₆H₈O      | 91.0546  | C₇H₇       | 107.0493 | C₅H₇O      |
| Cinchonine       | 6   | 294.1733  | 79.0544  | C₄H₇      | 184.0759 | C₁₂H₁₀NO   | 130.0654 | C₆H₈N      | 154.0653 | C₇H₇N      | 142.0654 | C₁₃H₈N      |
| Cotinine         | 1.4 | 177.1029  | 80.0499  | C₄H₈N     | 98.0606  | C₆H₈NO    | 146.0609 | C₆H₈NO     | 106.0657 | C₇H₇N      |
| Cylindropermopsin | 1.5 | 415.1166  | 336.1675 | C₁₅H₂₃N₃O₄ | 194.1293 | C₁₆H₁₆NO₃O₄ | 274.0864 | C₁₆H₁₆NO₃OS | 318.1570 | C₁₃H₂₈N₅O₈ |        |
| Kojic Acid       | 2.3 | 143.0336  | 113.0234 | C₅H₅O     | 126.0313 | C₆H₄O     | 97.02863 | C₅H₄O      | 87.00786 | C₅H₅O      |
| Microcystin LA   | 11.5| 910.4882  | 135.0808 | C₁₉H₁₁O    | 227.0224 | C₁₀H₈N₂O₅  | 299.0621 | C₁₂H₈N₄O     | 155.0689 | C₁₂H₈N₂O₆    | 297.0829 | C₁₃H₁₅N₄O₆ |
| Microcystin LF   | 12.3| 986.5225  | 135.0808 | C₁₉H₁₁O    | 213.0871 | C₁₀H₈N₂O₄  | 258.1855 | C₁₂H₈N₄O     | 461.2398 | C₁₂H₈N₄O₆    | 580.3016 | C₁₂H₈N₄O₇ |
| Microcystin LR   | 9.1 | 995.5375  | 135.0806 | C₁₉H₁₁O    | 382.2089 | C₁₀H₈N₂O₅  | 213.08728 | C₁₂H₈N₂O₅   | 470.2729 | C₁₂H₈N₄O₆   | 103.0544 | C₅H₇O      |
| Microcystin YR   | 11.7| 1045.5355 | 135.0806 | C₁₉H₁₁O    | 213.1364 | C₁₀H₈N₂O₆  | 265.1609 | C₁₂H₈O      | 323.1800 | C₁₂H₈N₄O₅   | 466.2589 | C₁₃H₈N₄O₄ |
| Microcystin LY   | 11.2| 1002.5353 | 135.0806 | C₁₉H₁₁O    | 375.1918 | C₁₀H₈N₂O₆  | 494.2616 | C₁₁H₈N₄O₅   | 213.08723 | C₁₂H₈N₂O₄   | 243.1343 | C₁₁H₈N₄O₅ |
| Nodularin        | 8.8 | 825.4505  | 135.080 | C₁₉H₁₁O    | 227.103 | C₁₀H₈N₂O₃  | 389.2074 | C₁₁H₈O₂N₂    | 691.3768 | C₁₂H₈N₂O₅    | 285.1668 | C₁₁H₈N₄O₆ |
| Ochratoxin-A     | 11.8| 404.0885  | 358.0835 | C₁₂H₁₆Cl₉O₄ | 257.0211 | C₁₁H₉ClO₄  | 239.0105 | C₁₂H₈ClO₄    | 120.0808 | C₁₃H₈N     | 211.0157 | C₁₃H₈ClO₃ |
| P-Coumaric acid  | 7.8 | 165.0544  | 91.0543 | C₇H₇      | 81.0336 | C₆H₅O     | 103.0542 | C₇H₇       | 119.0492 | C₆H₅O      | 147.0441 | C₇H₅O₂     |
| Scopolamine      | 6.21| 304.1538  | 138.0912 | C₁₉H₁₂NO   | 103.0542 | C₇H₇      | 110.09541 | C₁₅H₁₀N     | 103.0542 | C₆H₅O      | 121.0647 | C₆H₈O      |
| Thujone          | 12  | 153.1269  | 139.1120 | C₁₀H₅O    | 97.0650  | C₆H₅O     | 121.1043 | C₈H₃      | 109.0651 | C₇H₈O      | 135.1171 | C₁₀H₅      |
| Umbelliferone    | 8.12| 163.0386  | 107.0492 | C₇H₇O     | 95.0492  | C₆H₅O     | 91.0546  | C₇H₇      | 119.0493 | C₆H₅O      | 134.0363 | C₆H₆O₂     |
Table 5  
Results of the suspect screening with AIF acquisition in water samples.

| Comp No | Compound              | Molecular formula | \([M+H]^+\) | Transition 1 | Structure |
|---------|-----------------------|-------------------|-------------|-------------|-----------|
| 1       | Aspidospermine        | \(C_{22}H_{26}N_2O_2\) | 355.2373    | 119.0491    | \(C_7H_4O\) |
|         |                       |                   |             | 107.0491    | \(C_7H_4O\) |
| 2       | O-Acetyltropine       | \(C_{20}H_{27}NO_2\) | 184.1329    | 108.0807    | \(C_{11}H_{17}N\) |
|         |                       |                   |             | 109.0648    | \(C_7H_8O_2\) |
| 3       | Microcystin LR        | \(C_{10}H_{14}N_1O_{12}\) | 995.5545    | 135.0805    | \(C_9H_{11}O\) |
|         |                       |                   |             | 382.2089    | \(C_7H_{12}N_5\) |
| 4       | Heliotridine          | \(C_2H_4N\) | 156.1018    | 120.0808    | \(C\_B\_H\_N\) |
|         |                       |                   |             | 122.0965    | \(C_7H_4N\) |
| 5       | 4-Heptyloxybenzoic     | \(C_{14}H_{20}O\) | 237.1481    | 105.0699    | \(C_7H_11\) |
|         | acid                  |                   |             | 133.1012    | \(C_7H_11O\) |
| 6       | Hypoglicine A         | \(C\_H_2\_N_2\) | 142.0661    | 107.0492    | \(C\_H_2\_O\) |
|         | Salololn              |                   |             | 126.0550    | \(C\_H_2\_N_2O\) |
| 7       | Fumigaclavine C       | \(C_{22}H_3N_2O_2\) | 367.2374    | 105.0699    | \(C_7H_11\) |
|         |                       |                   |             | 119.0855    | \(C_7H_12\) |
| 8       | 4-hydroxymellin (\(R\)-reticuline   | \(C_{10}H_{13}O_4\) | 195.065    | 149.0234    | \(C_7H_11O_3\) |
|         |                       |                   |             | 121.0285    | \(C_7H_12O_2\) |
| 9       | Apio                   | \(C_{12}H_14O_4\) | 223.0961    | 105.0648    | \(C_7H_11O_3\) |
| 10      | Conhydrine             | \(C_2H_12N\) | 144.1382    | 107.0856    | \(C_7H_11O_3\) |
| 11      | 5-(N-Methyl-4,5-dihydro-1H-pyrrol-2-yl)pyrrolidin-2-ol | \(C_{10}H_{12}N_2O_2\) | 177.1019    | 120.0327    | \(C_7H_11O_3\) |
| 12      | Jervine                | \(C_{22}H_{27}O_1\) | 265.1432    | 163.0757    | \(C_7H_11O_2\) |
| 13      | Abietic acid           | \(C_{20}H_{28}O_3\) | 303.2387    | 257.2266    | \(C_7H_12O_2\) |
| 14      | Aspalatic acid         | \(C_{22}H_{28}O_5\) | 426.2999    | 187.1120    | \(C_7H_12O_2\) |
| 15      | Umbelliferone          | \(C_{12}H_13O_1\) | 163.0387    | 107.0492    | \(C_7H_11O\) |
| 16      | Swainsone              | \(C_{12}H_13O_1\) | 355.2372    | 270.1859    | \(C_7H_11O_3\) |
| 17      | Methoxycurarin         | \(C_{12}H_13O_2\) | 170.0545    | 91.0546     | \(C_7H_11O_2\) |
| 18      | Azelaic acid (Aspidomerine) | \(C_{12}H_13O_2\) | 225.1602    | 98.0604     | \(C_7H_11O_2\) |
| 19      | Coniferyl acetate      | \(C_{12}H_14O_2\) | 223.0961    | 91.0546     | \(C_7H_11O_2\) |
molecular ion m/z 181.07066 was searched. No D-glucose was found besides, the total absence of its precursor ptaquiloside brought to discard this compound as a tentative candidate. Finally, fragment analysis of 61 peaks revealed a strong similarity (1.0/1.0) with the 4-Heptyloxybenzoic acid [12] a carboxylic acid used with different purposes with no environmental importance for this work. Finally, 24 compounds have been detected and tentatively identified as suspect natural toxins. However, the confirmation to level 1 through mass spectrums comparison was carried out for 4 compounds with standards available (MC-LR, abietic acid, methoxycoumarin, and umbelliferone). Samples coming from the Czech Republic were collected in a blooming area which was characterized by green algal slime. This was the first signal to further investigate the presence of algal toxins such as microcystins. Here the tentative candidate microcystin LR was detected with the typical molecular ion at m/z = 995.5545. The doubly charged ion at 498.2822 m/z was also encountered at $T_r$ 9.12 min with the typical higher intensity respect to the molecular ion [13]. Finally, the MS/MS spectra revealed the presence of the typical fragment at 135.0803 m/z which is the exact mass of the ADDA fragment part of all the microcystins structure. After manual analysis of the MS/MS spectra, the precursor and 3 common fragments were found to be consistent with the MC-LR structure. Finally, the MC-LR was confirmed to level 1 using the standard solution that revealed the presence of the qualitative fragment ions in both mass spectrums. The same confirmation procedure was applied for methoxycoumarin, abietic acid, and umbelliferone while 20 structures were proposed as suspect natural toxins with an identification level 2.

Declaration of Competing Interest

The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

Acknowledgments

This project has received funding from the European Union’s Horizon 2020 research and innovation programme under the Marie Skłodowska-Curie grant agreement No. 722493 (NaToxAq); and by the Generalitat de Catalunya (Consolidated Research Group “2017 SGR 1404 - Water and Soil Quality Unit”). We also want to thank the Recetox colleagues Petra Laboha, Eliska Sychrova and Barbara Kubíčková for the sampling, logistics and supply of the water samples from Czech Republic.

Supplementary materials

Supplementary material associated with this article can be found, in the online version, at doi: 10.1016/j.mex.2021.101286.

References

[1] M. Picardo, J. Sanchis, O. Núñez, M. Farré, Suspect screening of natural toxins in surface and drinking water by high-performance liquid chromatography and high-resolution mass spectrometry, Chemosphere 261 (2020) 127888.
[2] B.F. Günthardt, J. Hollender, K. Hungerbühler, M. Scheringer, T.D. Bucheli, Comprehensive toxic plants-phytotoxins database and its application in assessing aquatic micropollution potential, J. Agric. Food Chem. 66 (29) (2018) 7577–7588.
[3] M. Krauss, H. Singer, J. Hollender, LC-high resolution MS in environmental analysis: from target screening to the identification of unknowns, Anal. Bioanal. Chem. 397 (3) (2010) 943–951.
[4] S. Vyse, H. Desmond, P.H. Huang, Advances in mass spectrometry based strategies to study receptor tyrosine kinases, IUCrJ 4 (2) (2017) 119–130.
[5] H. Horai, M. Arita, S. Kanaya, Y. Nihei, T. Ikeda, K. Suwa, Y. Ojima, K. Tanaka, S. Tanaka, K. Aoshima, Y. Oda, Y. Kakazu, M. Kusano, T. Tohge, F. Matsuura, Y. Sawada, M.Y. Hirai, H. Nakamichi, K. Ikeda, N. Akimoto, T. Maoka, H. Takahashi, T. Ara, N. Sakurai, H. Suzuki, D. Shibata, S. Neumann, T. Iida, K. Tanaka, K. Funatsu, F. Matsuura, T. Soga, R. Taguchi, K. Saito, T. Nishioka, MassBank: a public repository for sharing mass spectral data for life sciences, J. Mass Spectrom. 45 (7) (2010) 703–714.
[6] D.L. Lipschitz, W.C. Michel, Amino acid odorants stimulate microvillar sensory neurons, Chem. Senses 27 (3) (2002) 277–286.
[7] C. Rutkies, E.L. Schymanski, S. Wolf, J. Hollender, S. Neumann, MetFrag relaunched: incorporating strategies beyond in silico fragmentation, J. Cheminformatics 8 (1) (2016) 3.
[8] E.L. Schymanski, J. Jeon, R. Gulde, K. Fenner, M. Ruff, H.P. Singer, J. Hollender, Identifying small molecules via high resolution mass spectrometry: communicating confidence, Environ. Sci. Technol. 48 (4) (2014) 2097–2098.
[9] E. Sentandreu, M.D. Peris-Diaz, S.R. Sweeney, J. Chiou, N. Muñoz, S. Tiziani, A survey of orbitrap all ion fragmentation analysis assessed by an R metabolist package to study small-molecule metabolites, Chromatographia 81 (7) (2018) 981–994.

[10] X. Wang, N. Yu, J. Yang, L. Jin, H. Guo, W. Shi, X. Zhang, L. Yang, H. Yu, S. Wei, Suspect and non-target screening of pesticides and pharmaceuticals transformation products in wastewater using QTOF-MS, Environ. Int. 137 (2020) 105599.

[11] P.C.d.R. Aranha, L.H. Rasmussen, G.A. Wolf-Jäckel, H.M.E. Jensen, H.C.B. Hansen, C. Friis, Fate of ptaquiloside—a bracken fern toxin—in cattle, PLOS One 14 (6) (2019) 0218628.

[12] U.S. Environmental Protection Agency. CompTox Chemicals Dashboard. https://comptox.epa.gov/dashboard/DTXSID5036636 (accessed July 22, 2020), 4-Heptyloxybenzoic acid.

[13] D. Filatova, O. Núñez, M. Farré, Ultra-trace analysis of cyanotoxins by liquid chromatography coupled to high-resolution mass spectrometry, Toxins 12 (4) (2020) 247.