Assessment of Microbial Community Composition Changes in the Presence of Phytoplankton-Derived Exudates in Two Contrasting Areas from Chilean Patagonia

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Abstract: Patagonian fjords and channels in southern Chile are heterogeneous ecosystems characterized by the interaction of estuarine and marine waters influencing physical-chemical conditions and biological assemblages. Besides salinity, microbial communities from estuarine and marine origin are naturally subjected to changing organic matter quality and variable nutrient concentrations. In this study, we tackle the response of the bacterial community from estuarine and marine origins associated with two size classes (<0.7 µm and <1.6 µm) to the addition of sterile phytoplankton-derived exudates (PDE) compared to control conditions (no addition). Picoplanktonic cell abundance, active bacterial community composition analyzed through 16S rRNA sequencing, changes in dissolved organic carbon (DOC) and δ13C were determined over 5 and 15 days after PDE addition. Our results showed that the active marine bacteria were richer and more diverse than their estuarine counterparts, and were dominated by Alphaproteobacteria and Gammaproteobacteria, respectively. PDE addition in both the fractions and the sample origin resulted in an enrichment throughout the incubation of Rhodobacteracea and Cryimorphaceae families, whereas Epsilonproteobacteria (Arcobacteraceae) were mainly favored in marine bacteria. Picoplanktonic cell abundance increased with time, but higher cell numbers were found in PDE treatments in both size classes (>2 x 10^5 cell mL^-1). In all the experiments, DOC concentration decreased after eight days of incubation, but shifts in δ13C organic matter composition were greater in the estuarine experiments. Overall, our results indicate that despite their different origins (estuarine versus marine), microbial communities inhabiting the fjord responded to PDE with a faster effect on marine active bacteria.

Keywords: phytoplankton; microbial community composition; organic matter; Patagonian fjords

1. Introduction

The ecosystem englobing Patagonian fjords is one of the world’s largest estuarine environments, which supports high biological production and vertical carbon export fluxes [1–4]. The hydrographic structure of the fjords and channels of Patagonia is characterized by the presence of subantarctic water (SAAW) with high levels of nitrate and phosphate from the adjacent Pacific Ocean, and freshwater with high levels of silicic acid...
of terrestrial origin [5,6]. The surface freshwater layer that is formed by river discharges, precipitation, and glacier melting gradually mixes with the deeper and salty SAW through estuarine circulation. The influx of freshwater transports important loads of terrestrial organic and inorganic matter to the adjacent coastal areas [7,8]. Consequently, an exacerbated spatial and temporal variability in particulate and dissolved organic matter (POM and DOM) is an inherent characteristic of the fjords and channels of Chilean Patagonia [1–3]. This area is also characterized by high primary production due to the occurrence of phytoplankton blooms during spring [4,9]. All these characteristics influence autotrophic and heterotrophic microbial activity, while shaping the structure of the pelagic community [10].

In this dynamic and complex system, spatial and seasonal variability in microbial community composition have been linked to meltwater discharge associated to shifts in salinity, and macro and micronutrient availability [11–14]. Significant DOM release during phytoplankton blooms has also been associated with important changes in microbial diversity and metabolism, such as those observed in the Puyuhuapi Fjords [15]. Most of the DOM produced during phytoplankton blooms is first channeled through the microbial loop, where part of the DOM is assimilated to building up heterotrophic biomass, which is further transferred to higher trophic levels via predation, mobilizing a significant amount of carbon and energy that otherwise is not available at higher trophic levels [16]. Consequently, the interplay between microbial consumers and DOM release by primary consumers via extracellular release, cell breakdown (predation) or viral lysis is a key component of marine carbon cycling. The release of this DOM, which is usually considered highly labile and can be easily degraded by the heterotrophic bacterial community, can also enhance the remineralization of semi-labile and recalcitrant organic matter through the “priming effect” [17–20]. This process is especially relevant in estuarine systems, where pulses of labile organic matter derived from phytoplankton blooms might enhance land-derived carbon degradation, enabling the addition of new carbon sources to estuarine ecosystems [21].

It is known that the addition of labile DOM led to a higher relative abundance and a higher bacterial diversity of several bacterial groups, in particular Alphaproteobacteria, Bacteroidetes, and Verrucomicrobia (e.g., [22]). Under experimental conditions, Bacteroidetes and Alphaproteobacteria can respond under different DOMs derived from phytoplankton, responding to diatom cell lysis and cyanobacteria exudates, respectively [23]. Moreover, in areas where the priming effect is expected, such as river plumes, bacterial community composition did not change [24]. In this sense, most of the knowledge on priming effects comes from studies in soil; the mechanisms controlling the pool of recalcitrant DOM in oceans and, particularly, in fjords are less understood [25].

Herein, we study the differential responses of estuarine and marine bacterial communities to the introduction of fresh DOM derived from phytoplankton and hypothesize that the expected “priming effect” could favor the activity of specific heterotrophic bacteria inhabiting estuarine areas usually exposed to more recalcitrant organic matter. Using an experimental approach, we evaluated changes in the abundance, composition, and diversity of bacterioplankton communities as a response to phytoplankton-derived exudates (PDE; labile DOM).

2. Materials and Methods

2.1. Environmental Conditions

The study area is located in the Kruger Channel (station 5, 47.89° S–74.56° W) and in the middle of Martinez Channel near the mouth of the Steffen fjord (station 14, 47.81° S–74.75° W; Figure 1) in Chilean Patagonia. This sampling area is under the influence of freshwater discharge from the two largest rivers in Chile (Baker and Pascua rivers). Sampling was conducted during the austral spring of 2014 (October) under the framework of CIMAR20-FIORDOS expedition. Two contrasting stations were selected, station 5 (marine influence) and station 14 (estuarine influence). Experiments were carried out on October 4th and 14th. Water column temperature and salinity were obtained using a CTD
Discrete water samples were collected using Niskin bottles (2.7 L) arranged on a CTD rosette at standard depths (0, 5, and 25 m). At each depth, subsamples were collected in duplicate for the determination of macronutrients (silicic acid, nitrate and orthophosphate) and chlorophyll-a. Nutrient samples (11 mL) were filtered through 0.7 µm filters (GF/F; Whatman, Maidstone, UK), placed in 11 mL HDPE plastic tubes (acid washed), and stored at −20 °C prior to laboratory analysis. Nutrients were determined using standard colorimetric techniques [26]. These analyses were performed using a Seal analytical AutoAnalyzer AA3 segmented flow at Biochemistry laboratory at Universidad de Concepción, Chile. Chlorophyll-a was measured in duplicate with the fluorometric technique by filtering 200 mL of seawater through 0.7 µm GF/F filters (Parson et al., 1984).

Figure 1. Map showing the location of the stations sampled (dots) in October 2014 in northern Chilean Patagonia.

2.2. Experimental Design

To study the response of natural marine and estuarine microbial assemblages to phytoplankton-derived exudates (PDE), surface water (~12 L) from a 2 m depth was collected at each station for controls and PDE treatments and the two size fractions were studied. A total of 24 bottles per experiment were used (500 mL, Duran Schott sterile bottles). The experimental incubations consisted of two size fractions to favor microbial communities and avoid grazers; the first was 1.6 µm filtered water (APFA04700, Millipore, Burlington, MA, USA), and the second consisted of 0.7 µm filtered water (GF/F, Whatman). The natural bacterial community within the 0.7 and 1.6 µm size class, hereafter named as “controls”, was collected at each station. The DOM experiments consisted of six glass bottles per size fraction that were enriched with 0.5 mg L⁻¹ of DOC from phytoplankton-derived exudates. The DOM stock solution was obtained from phytoplankton exudates derived from cultures of Chaetoceros spp. and Thalassiosira spp., two dominant species in southern Chile, at approximately 4 × 10⁶ cells/mL densities. Cultures of Chaetoceros spp. and Thalassiosira spp. were obtained from FICOLAB laboratory at the University of Concepción. A liquid cell suspension from cultures was gently filtered through 0.2 µm, avoiding potential contamination of the samples.
All the bottles were incubated in the dark and in situ simulated temperature conditions using an onboard marine water recirculation system. The recirculation system consisted of two large incubators connected by hoses, through which surface seawater from each station circulated. Subsamples were collected for DOC, δ\(^{13}\)C, picoplankton abundance and nanoplanlton abundance, and to determine the active bacterial composition at intervals of 0 (initial condition), 3, 5, 8, 12, and 15 days after incubation. Furthermore, nitrate (NO\(_3^-\)), silicic acid (Si(OH)\(_4\)) and orthophosphate (PO\(_4^{3-}\)) were determined at sampling intervals of 0, 5, and 15 days using the same analytic procedures as described before. Inorganic nutrient results are available as Supplementary Materials.

A sample of 500 mL was filtered through a combusted 0.7 µm GF/F filter for particulate organic carbon analysis. The carbon (δ\(^{13}\)C) isotope was analyzed by mass spectrometry using a Thermo Scientific Delta V Advantage IRMS with EA-2000 Flash Elemental analyzer at the Laboratory of Biogeochemistry and Applied Stable Isotopes (Pontificia Universidad Católica, Santiago, Chile). The δ\(^{13}\)C (‰) standard used was Pee Dee Belemnite (PDB). For DOC determinations, a subsample of 40 mL was filtered through a 0.22 µm sterile syringe MilliporeMilli@ (Merck Millipore, Burlington, MA, USA), stored in 40 mL amber I-Chem 200 vials (Thermo Scientific CertifiedTM, Waltham, MA, USA), and acidified with 300 µL trace metal grade HCl. DOC was analyzed using the high-temperature catalytic oxidation (HTCO) technique [27] using a Shimadzu TOC-5000 carbon analyzer. Picoplankton abundance was analyzed by flow cytometry [28] with a FACSCalibur flow cytometer (Becton Dickinson, Franklin Lakes, NJ, USA), and nanoplanlton abundance was analyzed by epifluorescence microscopy [29].

2.3. Active Bacteria, Nucleic Acid Extraction, and Sequencing Procedure

The bacterial community at each time interval was concentrated from a 100 mL water sample onto 0.22 µm hydrophilic PVDF filters (Millipore) using a sterilized syringe and a 25 mm filter holder (Merck Millipore, Burlington, MA, USA). The filters were preserved in 1.5 mL cryovials with RNAlater solution (Ambion, Austin, TX, USA), frozen in liquid nitrogen in the field and stored at −80 °C prior to RNA extraction in the laboratory. RNA samples were taken in duplicate and were extracted using a Mirvana kit (AM1560; Ambion, Austin, TX, USA), as described previously by Valdès et al. [30]. RNA extracts were then treated to remove DNA tracer with a TURBO DNA-free kitTM (Ambion) and quantified using a Qubit fluorometer (Thermo Fisher Scientific), following the manufacturer’s instructions. The RNA was reverse-transcribed to cDNA using random primers with the ImProm-II Reverse Transcription System (Promega, Madison, WI, USA). Amplification was checked using standard PCR to amplify the V1–V3 region of bacterial 16S rRNA (27F-519R). The cDNA templates were sequenced using the Illumina high-throughput sequencing method at the Research and Testing Laboratory (RTL, Lubbock, TX, USA).

2.4. 16S rRNA Gene Sequence Analyses and Microbial Community Characterization

The 16S rRNA gene sequence analyses were performed using Mothur software v1.35.1 [31]. Sequencing data sets were curated by quality filtration to minimize the effects of random sequencing errors by eliminating short sequence reads <200 bp, sequences that contained more than one undetermined nucleotide, and sequences with a maximum homopolymer length of eight nucleotides. Thereafter, chimeric sequences were identified using the chimera UCHIME algorithm and removed [32]. The curated 16S rRNA gene sequences were taxonomically classified using automatic software pipeline SILVAngs available from https://www.arb-silva.de/ (accessed on 14 December 2021) [33]. At the marine station, a total of 143,842 gene sequences were analyzed, ranging between 7861 and 21,752 sequences for each library. At the estuarine station, 127,158 sequences were analyzed from 4384 and 11,212 for each library. Libraries were deposited in the European Nucleotide Archive (ENA) under study accession PRJEB42326 with the following run access numbers: ERS5801012–ERS5801039.
The composition was identified at phyla, class, order, and family for taxonomic levels as abundant (>0.5%) and others (<0.5%) in total sequences retrieved from each library. Since the number of sequences per sample was variable, we previously rarefied the different library sizes using a subsampling routine in Mothur software [31] to reflect the lowest number of sequences encountered (station 5: 7861 and station 14: 4384).

2.5. Statistical Analysis

Statistical analysis of the effect of the different treatments (PDE addition, incubation time and different sizes of bacterial assemblages) on DOC, \( \delta^{13}C \), picoplankton abundance, and Shannon index, were performed using a 3-way analysis of variance (ANOVA) after checking for normality assumptions (Kolmogorov–Smirnov test) and homoscedasticity (Levene’s test). A pairwise multiple comparison was performed using a Tukey test as a posteriori analysis.

Bacterial community structures were compared using ordination Bray–Curtis similarities, visualized in Unweighted Pair Group Method with Arithmetic mean (UPGMA) dendrograms, whose nodes were further tested using a bootstrap analysis. A multivariate analysis was also used to analyze the variability of biological and environmental variables, using PRIMER v.6 and the add-on PERMANOVA+ software package. The permutational multivariate analysis of variance (PERMANOVA) with a fixed factor was used to investigate the differences in bacterial community composition for treatments with PDE addition and control, treatment with different bacterioplankton sizes, and between the different incubation times, in both experiments. Principal coordinate analysis (PCoA) was performed to visualize the patterns of the operational taxonomic units (OTUs) of the bacterial community in response to different treatments.

3. Results
3.1. Hydrographic and Biogeochemical Conditions

The surface layer was generally characterized by low temperature, 8–9.5 °C, and variable salinity considering the east-to-west transect sampled (Figure 2A,B). Marine station presented the lowest temperature (~8 °C) at the surface up to 15 m depth with a weak thermocline, while the estuarine station presented a subsurface minimum between 5 and 10 m, causing a thermal inversion from station 14 to station 5, as a result of the freshwater inflow towards the west (Figure 2A). As expected, salinity controls the vertical structure along the east to west transect, with lower salinities close to the outlet of Steffen fjord with a marked halocline at 8 m at station 14 (7.5 and 22 PSU), followed by a gradual increase in salinity towards station 5 (32 PSU). The estuarine waters associated with the Steffen fjord were characterized by lower nitrate and phosphate (below 7.5 and 0.5 μmol L\(^{-1}\), respectively; Figure 2C–E) and higher silicate concentrations (15–25 μmol L\(^{-1}\)) compared with the Kruger channel. At this site and below the pycnocline, the phosphate and nitrate concentrations were higher (15–26 and 1–2.5 μmol L\(^{-1}\), respectively), while the silicate concentrations remained <7.5 μmol L\(^{-1}\) compared to the Steffen fjord station (Figure 2).
Figure 2. (A) temperature (°C), (B) salinity (PSU), (C) nitrate (µmol L⁻¹), (D) phosphate (µmol L⁻¹), and (E) silicate (µmol L⁻¹) spatial variability in an east-(Kruger channel)-to-west (Steffen fjord) transect. Surface water for the experiments was obtained from the marine-influenced station 5 and the estuarine station 14.

The chlorophyll-a (Chl-a) concentrations showed higher concentrations at surface at a 0 m depth for total Chl-a and <20 µm, but for both <20 µm and 2–20 µm sizes, the maxima occurred at 5 m at the marine station 5 (Table 1). At the estuarine station 14, higher values of chlorophyll-a were observed at subsurface depth (5 m) for total and almost all fractions, except for the 2–20 µm size fraction, where the highest concentration was observed at 10 m depth.

Table 1. Fractionated chlorophyll-a concentrations (mg m⁻³) in marine station 5 and estuarine station 14.

| Station | Depth (m) | Total Chlorophyll-a (mg m⁻³) | <2 µm | 2–20 µm | >20 µm |
|---------|-----------|-------------------------------|-------|---------|--------|
|         |           |                               |       |         |        |
| 5       | 0         | 8.38 ± 9.34                   | 0.74 ± 0.00 | 1.64 ± 0.00 | 5.99 ± 9.34 |
|         | 5         | 4.26 ± 0.91                   | 0.93 ± 0.01 | 2.57 ± 1.26 | 0.76 ± 2.18 |
|         | 10        | 3.06 ± 0.17                   | 0.28 ± 0.03 | 0.74 ± 0.03 | 2.03 ± 0.16 |
|         | 25        | 0.98 ± 0.03                   | 0.08 ± 0.01 | 0.30 ± 0.11 | 0.60 ± 0.15 |
|         | 50        | 0.10 ± 0.02                   | 0.24 ± 0.00 | 1.38 ± 0.00 | 0.00 ± 0.00 |
| 14      | 0         | 1.40                          | 0.18   | 1.02    | 0.20   |
|         | 5         | 10.77                         | 0.61   | 1.83    | 8.33   |
|         | 10        | 2.61                          | 0.37   | 2.39    | 0      |
|         | 25        | 0.11                          | 0.12   | 1.11    | 0      |
|         | 50        | 0.19                          | 0.06   | 0.62    | 0      |

3.2. Changes in Biogeochemical Conditions after PDE Addition to the Different Picoplanktonic Fractions

The PDE enrichment and control experiments resulted in DOC changes over time. At the marine station, DOC concentration was significantly higher in the treatment with the PDE addition compared to the controls (Figure 3B; ANOVA; p < 0.0001) but no significant differences were observed between size classes (ANOVA; p = 0.7). Furthermore,
slight differences were observed between times, highlighting a slight increase at 5 days of incubation in the controls compared to the initial time (ANOVA and Tukey's test; \( p < 0.05 \)). At the estuarine station (Figure 3D), the DOC concentration was significantly higher with PDE addition (ANOVA, \( p < 0.0001 \)), but no significant differences were found between the different size classes of natural bacterial assemblages or between incubation times (ANOVA, \( p = 0.22 \) and \( p = 0.08 \), respectively).

### Figure 3

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\begin{align*}
\delta^{13}C \text{ (‰)} \quad (A, C) \quad \text{and DOC concentration (gr m}^{-3}) \quad (B, D) \text{ from the different treatments, with PDE addition and control, and for the different bacterioplankton. Left panel corresponds to marine station 5 and the right panel to estuarine station 14.}
\end{align*}
\]

Along the course of the experiment, significant differences were observed in \( \delta^{13}C \) between the treatments with PDE addition and controls at the marine station (ANOVA; \( p < 0.0001 \)) and the estuarine station (ANOVA; \( p = 0.015 \); Figure 3A,C). At marine station 5, the higher enrichment values were found in the controls (\(-25\% \text{ in average}\)) and the lowest values were found in the treatments with PDE addition (\(-28\% \text{ in average}\)). A significant enrichment was observed in the first 5 days of incubation in the treatments with PDE addition compared to the controls, followed by a decrease after 8 and 12 days in the treatment with 1.6 \( \mu \)m and 0.7 \( \mu \)m size fraction classes, respectively (ANOVA and Tukey’s test; \( p < 0.05 \)). However, non-significant differences were found between the different size classes of bacterial assemblages (ANOVA, \( p = 0.17 \)) at station 5. At station 14, no significant differences were observed between the times (ANOVA, \( p = 0.06 \)) or between the different cell size classes (ANOVA, \( p = 0.4 \)).

### 3.3. Changes in Picoplankton Abundance during the Incubations

The changes in picoplankton abundance are shown in Figure 4. During the experiment at the marine station (Figure 4A), the picoplankton abundances were statistically different between the treatment with PDE addition and the control (ANOVA, \( p < 0.0001 \)). Similarly, the initial picoplankton abundances were statistically higher than between days 5 and 8 (ANOVA and Tukey’s test; \( p < 0.05 \)). By contrast, no significant differences were observed between the different size classes of picoplankton assemblages (ANOVA, \( p = 0.202 \)). However, the interaction of the factors (PDE additions per size classes of bacteria and per
time) showed significant effects (ANOVA, \( p < 0.0001 \)). The picoplankton abundances at the Steffen fjord station showed significant differences between the different picoplankton size classes (0.7 µm and 1.6 µm; ANOVA, \( p = 0.01 \)). In general, the treatment inoculated with the larger community (1.6 µm) showed higher abundances than the lowest size class (0.7 µm). The treatment inoculated with the smallest picoplankton community (0.7 µm) increased their abundance from the first and third day of incubation in the treatment with PDE addition and the control, respectively. In particular, this increment was abrupt for the treatment with PDE additions at day 3 (Figure 4B). Meanwhile, the treatment inoculated with the largest picoplankton decreased after the first day of incubation. However, no significant difference was observed between the treatments with PDE addition and the controls (ANOVA, \( p = 0.45 \)), nor between the different times of incubation (ANOVA, \( p = 0.48 \)), but the interactions between the factors showed significant effects (ANOVA, \( p = 0.002 \)).

Figure 4. Non-fluorescent picoplankton abundance (10^3 cell mL\(^{-1}\)) from the different treatments, with PDE addition and controls, and for the different size fraction of natural bacterial assemblages (0.7 µm and 1.6 µm). The upper panel corresponds to marine station 5 (A) and the lower panel to estuarine station 14 (B).

3.4. Active Bacterial Community Structure and Composition Response to PDE Addition

The alpha diversity derived from the normalized 16S rRNA libraries (similar sequence numbers) for comparison is shown in Figure 5A,B (station 5) and Figure 5C,D (station 14). At marine station 5, the bacterial taxonomic richness determined (OTU number) and expected on the chao1 index was higher at the initial times in the controls (without PDE addition) in both bacterial size fractions (Figure 5A,B). After PDE addition, the determined and expected richness decreased in the treatment incubated with the smaller picoplankton size class (Figure 5A) and increased in the treatment inoculated with the larger size fraction (Figure 5B). In accordance with this, the Shannon diversity was statistically higher at the initial time in the control in both bacterial size fractions. Furthermore, the Shannon index was statistically different between the controls and PDE additions (ANOVA, \( p = 0.008 \)), between the different picoplankton size classes (ANOVA, \( p = 0.02 \)), and between the incubation times (ANOVA, \( p = 0.005 \)). At the estuarine station (station 14), the treatment
incubated with the smaller bacterial size fraction showed a decrease in the expected and determined richness in the controls, as well for the treatment with PDE addition (Figure 5C), by contrast, the Evenness index increased throughout the incubation. Slight differences were observed in the determined and expected richness in the treatment incubated with the larger size fraction (1.6 µm), as well as in the Shannon and Evenness indexes in the treatment with PDE addition and the control. Accordingly, non-significant differences were observed in the Shannon index between PDE addition and the control (ANOVA, \( p = 0.62 \)), either in the different sizes of picoplankton community (ANOVA, \( p = 0.26 \)) or between the times (ANOVA, \( p = 0.47 \)).

Figure 5. Richness, chao1, diversity (Shannon), and evenness in the bacterial community and over the time course of the experiment for the different treatments, with PDE addition and controls, and for the different bacterial size fractions, 0.7 µm (A,C) and 1.6 µm(B,D). (Left panel) corresponds to marine station and (Right panel) to estuarine station.

At marine station 5, the initial bacterial community composition (control T0; Figure 6) for the treatment incubated with the lower size fraction (0.7 µm) was characterized by a higher contribution of Alphaproteobacteria, mainly SAR11 (43% average between replicates) and Gammaproteobacteria (33% in average of libraries), mainly SAR86 clade. Similarly, the treatment incubated with the 1.6 µm picoplankton size class was characterized by a higher contribution by Alphaproteobacteria (mainly SAR11 and Rhodobacterales), reaching 45% of the libraries, and Gammaproteobacteria (SAR86 clade), with 29% (on average). At estuarine station 14, close to Steffen fjord, the initial communities (control) of the treatment incubated with the 0.7 µm size class was characterized by a higher contribution of Gammaproteobacteria (45% average between replicates), mainly Betaproteobacteria, SAR86 clade, and Oceanospirillales, followed by Alphaproteobacteria (42% average), mainly SAR11 clade. The PDE addition treatments were characterized by Epsilonbacteraeota and Bacteroidetes (mainly Flavobacteria).
Figure 6. Similarity dendrograms based on the 16S rRNA and relative contribution of the abundant bacteria phyla and Proteobacteria subclasses for both stations and all treatments and incubation times. Clustering was on the basis of a distance matrix computed using the Bray–Curtis index of similarity. The dendrogram was inferred with the unweighted pair-group average algorithm (UPGMA). Bacterial communities in the samples connected with red branch lines are not significantly different (SIMPROF test, $p < 0.05$).

A clustering analysis using bacterial orders showed that the variability was mainly associated with the PDE additions and incubation times (Figure 6). In general, the SIMPER analysis indicated that 37 families accounted for 50% of the dissimilarities between the enrichment treatment (PDE additions) and the controls, and between the different picoplankton sizes (Figure 7). At the marine station (Figure 7), the large picoplankton size class (1.6 $\mu$m) treatment presented an enrichment of Saccarospirillaceae, from 0.9% to 34%, from day 5 until the end of the incubation (day 15). A similar enrichment was observed for Colwelliaceae (from 0.4 to 16%) and Nitrincolaceae (from 0.03 to 14%; especially in the incubation day 5). By contrast, families such as Thioglobaceae, SAR86 clade, Burkholderiaceae, and SAR202 clade, among others, reduced their contribution. After PDE addition, Rhodobacteraceae increased its contribution from 11.5 to 50.8% at the end of incubation, followed by Arcobacteraceae (from 0.4 to 14%), Marinobacteraceae (from 0.01 to 14%), Marinomonadaceae (from 0 to 10%), and Parvibaculaceae (from 0.2 to 13%). This was mainly observed at incubation day 5. In the control treatment associated with the smaller picoplankton community (0.7 $\mu$m), a greater relative abundance of Rhodobacteraceae (from 0.7 to 15%), and Burkholderiaceae (from 0.2 to 34%) were observed at incubation day 5. After 15 days of incubation, a higher contribution of Saccarospirillaceae (from 0.3 to 43%) and Nitrincolaceae (from 0.02 to 18%) was observed. After PDE addition, a higher increase in the abundance of Rhodobacteraceae (from 3 to 45%), Marinobacteraceae (from 0 to 35%), and Xanthomonadaceae (from 0 to 17%) was observed in the sequencing libraries between 5 and 15 days of incubation. Furthermore, a slight increase in Obscuribacterales, Arcobacteraceae, Cryomorphaceae, Parviculaceae, and Alteromonadaceae was observed. The results from the main PERMANOVA test revealed significant differences between the incubation times (pseudo-$F = 12.7$, $p = 0.005$), as well as between the treatment with PDE addition and the control (pseudo-$F = 9.26$, $p = 0.013$), and non-significant differences were
observed between the different sizes of bacterioplankton assemblage (pseudo-F = 4.05, \( p = 0.09 \)).

Figure 7. Contribution of active families through the incubation based on the SIMPER results. List of the OTUs explaining 50\% of the dissimilarity observed during the experiment between treatments with PDE addition and control (without PDE) for each station. Upper panels correspond to marine station 5 and lower panels to estuarine station 14.

At estuarine station 14 (Figure 7), in the large size fraction treatment (1.6 \( \mu m \)), the SIMPER analysis revealed an enrichment in the relative abundance of Arcobacteraceae towards the end of incubation (from 0.1 to 56\%). A notable reduction in Saccharospirillaceae relative abundance in the sequencing libraries was observed during the incubation (from 64\% to 0.02\%); however, this family was not observed in both replicates at the initial time (T0). The dissimilarities between the incubation with PDE addition and the controls were explained by an increase in the relative abundance of Rhodobacteraceae at the initial time, followed by an increase of Cryomorphaceae (from 4 to 20\%) and Marinobacteraceae (from 10 to 20\%) at incubation day 5. However, towards incubation day 15, an Arcobacteraceae enrichment from 0.02 to 69\% was observed, along with a decrease in the relative abundance of Rhodobacteraceae, Cryomorphaceae, and Marinobacteraceae. The smaller size fraction (0.7 \( \mu m \)) control treatment was characterized by an increase of Arcobacteraceae (from 0.1 to 56\%), Flavobacteriaceae (0 to 6\%), and Burkholderiaceae (from 15 to 37\%) between the 5th and 15th day of incubation. Once the addition of PDE occurred, an increase in Rhodobacteraceae, Marinobacteraceae, and Cryomorphaceae was observed during all the times compared to the control treatments. The Arcobacteraceae (from 0.02 to 69\%) abruptly increased at day 15, followed by an increase in Chloroplast and Pseudohongiellaceae at the incubation day 5 and towards the end of the incubation, respectively. However, the main PERMANOVA test showed non-significant differences between the treatment with PDE additions and the control (pseudo-F = 2.76, \( p = 0.066 \)) and
between the different size classes (pseudo-F = 2.18, p = 0.1), but significant differences were observed between the incubation times (pseudo-F = 2.9, p = 0.03).

Furthermore, significant differences were observed between the different stations, (PERMANOVA; pseudo-F = 2.65, p = 0.001). Marine station 5 was characterized by a higher contribution of Alphaproteobacteria than estuarine station 14, which was characterized by the contribution of Gammaproteobacteria. The main dissimilarities between the stations were attributable to the contribution of Betaproteobacterales, Oceanospirillales, Campylobacterales, Rhodobacterales, and Alteromonadales, accounting for 30% of the dissimilarities between both stations.

A principal coordinate analysis (PCoA; Figure 8) considering the order taxa contribution and environmental parameters indicated that the marine station (Figure 8A), samples with PDE addition were associated with Si and δ13C. At the estuarine station (Figure 8B), the samples were mostly grouped by incubation time rather than PDE addition; the initial time was associated with the Si concentrations.

Figure 8. Principal component ordination of treatments with PDE and control based on Bray–Curtis similarity at order taxonomic level. (A) marine station 5 and (B) estuarine station 14. Vectors indicate the best environmental variables (normalized transformed) correlated with ordinations and vector lengths correspond to the correlation values.

4. Discussion

DOMs derived from phytoplankton reactivity and their use by fjord microbes were assessed by analyzing the response of bacterial communities in surface waters to PDE addition from contrasting environments with freshwater and marine influences along a land–ocean transect in Patagonian fjords. Our results indicated an increased bacterial abundance and changes in the community composition as immediate responses to PDE addition. The hydrography and nutrient concentrations registered in the water column during our study showed the influence of freshwater and marine conditions, as can be expected for this kind of ecosystem in Patagonian fjords during the austral spring [2]. The thicker and more extensive lower-salinity layer at the surface indicated that the fjord is subjected to freshwater discharge in spring from Steffen fjord, resulting in a strong land–ocean gradient, as has been reported by other authors [2,6,8]. The surface estuarine waters were characterized by lower nitrate and phosphate, but high silicate, compared to the marine waters. These characteristics in fjord waters play a key role in influencing the phytoplankton biomass and structure [2,34], as well as the microbial community structure and dynamics in Chilean fjords [15].

During our study, the active bacterial community composition suggests differences in the surface water between the marine and the estuarine types, with the dominance of Alphaproteobacteria and Gammaproteobacteria, respectively. These bacterial taxa were
reported to be highly abundant in surface waters of high latitudes and fjord ecosystems in DNA survey studies [11–13,35]. Bacteroidetes, Alphaproteobacteria, and Betaproteobacteria have been reported in freshwater and marine waters from Arctic fjords, and the differences between both microbial communities are mainly due to the presence of Epsilonproteobacteria in marine waters and of Gammaproteobacteria, Acidobacteria, Firmicutes, and Verrucomicrobia in estuarine waters [11,12,36]. In Chilean Patagonia and Antarctic waters at higher taxonomic levels, the dominance of Bacteroidetes and Alphaproteobacteria members [15,35] has been observed. In our study, the Rhodobacteraceae, SAR11 clade I, and SAR86 families were characteristic of the marine waters, and in the estuarine freshwater, the presence of Rhodobacteraceae, Burkholderiaceae, SAR86 clade, and SAR11 clades I and III were noted at the beginning of the incubation. Rhodobacteraceae and SAR11 clade I have been described as a representative family associated with estuarine freshwaters in Patagonian fjords [15,35]. Although the taxa of the active bacteria coincide with those of previous studies, bacterial community changes could be expected considering prefiltration and the time to set up the beginning of the experiments (~1 h). Water collection and experimental incubations have been reported to trigger significant changes in the microbial community transcription associated, for example, with Alphaproteobacteria (SAR11-like) and Gammaproteobacteria (Collwelia from Alteromonadales) in the Eastern South Pacific [37].

4.1. Potential Organic Matter Degradation by Different Small Microbial Size Fractions

Phytoplankton and heterotrophic prokaryotes are major components of the microbial food web and interact continuously in very complex ways. The mainly fresh phytoplankton-derived DOM from phytoplankton blooms is a major energy source driving heterotrophic prokaryote and phytoplankton communities [20]. In accordance with this, the experimental incubations with PDE amendments in the marine and estuarine waters influenced DOM degradation, as evidenced by the reduction of \( \sim 1.5 \) mg L\(^{-1}\) of DOC throughout the incubation, especially after 3 days of incubation in the estuarine water experiment. The DOC decrease was characterized by a major picoplankton bloom (Figure 5B), suggesting a net effect of DOC consumption by the heterotrophic bacterioplankton community. Moreover, the PDE amendments showed nitrate and phosphate accumulation at the end of the incubation in the marine experiment and during the entire incubation at the estuarine station (Supplementary Material Figure S1), indicating the presence of remineralization end-products. These results support previous studies suggesting the higher nitrification potential in surface waters and after organic-matter-addition experiments, such as the use of dissolved salmon food pellets [38].

The addition of PDE to natural bacterioplankton assemblages generated changes in abundance. Our results showed significant differences in bacterioplankton abundance between the treatment with PDE addition and controls only in the marine station. However, the addition of PDE did not show a higher increase in bacterial abundance compared to the control, and, in some cases, the control had greater abundances over time (e.g., Figure 4A, control 1.6 \( \mu \)m). In any case, the microbial response was more evident at the initial times, between 2 and 5 days of incubation. Furthermore, the presence of heterotrophic nanoflagellates (Supplementary Material Figure S2) can explain picoplankton decrease by predation at the marine station. At the estuarine station, an intense bloom after 5 days of incubation in the treatment with PDE addition and inoculation with the small bacterioplankton size fraction was observed (Figure 4B). This bloom was also associated with a decrease in DOC concentration (Figure 3D). Many studies have shown the response of bacterioplankton to labile organic matter addition, as in our PDE treatments, as a driver of bacterioplankton growth, highlighting a rapid response (hours) in the form of bacterioplankton abundances after the addition of labile organic matter, independently of their origin [22,39].
4.2. Active Bacterial Community Structure and Composition Responses during Surface Marine and Estuarine Water Incubations

The bacterial community structure based on rarefying diversity evidenced differences between the control and PDE addition. In general, alpha diversity increased after the PDE additions compared to the initial time (particularly at the middle times). Proteobacteria, Bacteroidetes, and Epsilonbacteraeota phyla presented a greater contribution to our libraries and were the phyla showing the most significant differences during the experiments. The positive PDE taxon-specific active bacterial responses were attributed to Alphaproteobacteria, Gammaproteobacteria, and Epsilonbacteraeota, particularly Rhodobacteraceae, Marinobacteraceae (Marinobacter), Arcobacteraceae, and Cryomorphaceae regardless of the bacterioplankton community sizes (1.6 or 0.7 µm). The interpretation of the diversity and structure changes during the microcosm incubation is always a controversial issue due to a potential confinement effect. As in previous studies, an increase in Gammaproteobacteria taxa, probably related to the confinement [38,40], occurred during our experiments. However, the bottle effect usually results in a decrease in total microbial diversity [41]; this was only detected at station 5 for the treatment incubated with the larger microbial size (Figure 5).

Recently, Kieft et al. [23] demonstrated that DOM composition has strong bottom-up control over the assembly of the active subset of the total community, which is generally composed of populations with specialized metabolic and ecophysiological traits, facilitating the efficient turnover of the highly heterogeneous pool of DOMs available throughout a phytoplankton bloom. In this sense, the increase in Gammaproteobacteria has been widely associated with phytoplankton bloom dynamics and with an increased concentration of DOMs in lower latitudes [42,43]. Teeling et al. [43] indicate that distinct populations of Bacteroidetes, Gammaproteobacteria, and Alphaproteobacteria are specialized to use algal-derived organic matter, providing ecological niches in which specialized populations could bloom. In polar areas, changes in bacterial diversity have been observed after increases in chlorophyll-a concentrations during summer blooms, where Gammaproteobacteria, Alphaproteobacteria, and Bacteroidetes have the largest relative abundance and activity [44,45]. The Rhodobacteraceae family was found in all the treatments with PDE and controls; although this family was not abundant in the initial communities in this study, they become dominant members of fjords microbial communities during times of high levels of DOMs derived from phytoplankton inputs. The Rhodobacteraceae family has been detected in estuarine fresh waters in the Puyuhuapi fjord in Chilean Patagonia and is positively correlated with chlorophyll-a concentration [15,35]. Furthermore, this taxon is a specialized degrader of phytoplankton-derived compounds, such as DOM [46].

The Cryomorphaceae family (Bacteroidetes) showed lower abundances in the controls, and was even undetected in some cases, but increased their abundance significantly after PDE addition. This family has been positively correlated with phytoplankton community composition, especially with dinoflagellates, in Chilean Patagonia fjords [15]. Bacteroidetes are among the most abundant phyla in coastal areas [47]; they can degrade complex organic matter and are specialized in the degradation of biopolymers, such as proteins [43,48]. In this sense, the exacerbated input of nutrients that results from salmon culture and the increase in marine freshwater temperature and stratification promote shifts in phytoplankton community composition and blooming phenology [49], thereby modifying the microbial community structure and selecting the best-adapted taxa, even rare taxa, which was observed in Cryomorphaceae family during our experiments. Furthermore, modifying the activity and community structure of microbes could significantly affect the cycling of organic matter in Patagonian fjords. Altogether, we conclude that the organic matter additions derived from phytoplankton exudates can significantly modify the structure of fjord microbial communities, favoring groups adapted to rapidly react to changes in their organic matter availability, which allows them to outcompete other abundant groups in the field during marine phytoplankton blooms.
Supplementary Materials: The following are available online at https://www.mdpi.com/article/10.3390/d14030195/s1. Figure S1: Nutrient dynamic in response to PDE addition. Figure S2: Autotrophic and heterotrophic nanoflagellates during incubation experiments.

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