Response of soil respiration and ecosystem carbon budget to vegetation removal in Eucalyptus plantations with contrasting ages

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Reforested plantations have substantial effects on terrestrial carbon cycling due to their large coverage area. Although understory plants are important components of reforested plantations, their effects on ecosystem carbon dynamics remain unclear. This study was designed to investigate the effects of vegetation removal/understory removal and tree girdling on soil respiration and ecosystem carbon dynamics in Eucalyptus plantations of South China with contrasting ages (2 and 24 years old). We conducted a field manipulation experiment from 2008 to 2009. Understory removal reduced soil respiration in both plantations, whereas tree girdling decreased soil respiration only in the 2-year-old plantations. The net ecosystem production was approximately three times greater in the 2-year-old plantations (13.4 t C ha⁻¹ yr⁻¹) than in the 24-year-old plantations (4.2 t C ha⁻¹ yr⁻¹). The biomass increase of understory plants was 12.6 t ha⁻¹ yr⁻¹ in the 2-year-old plantations and 2.9 t ha⁻¹ yr⁻¹ in the 24-year-old plantations, accounting for 33.9% and 14.1% of the net primary production, respectively. Our findings confirm the ecological importance of understory plants in subtropical plantations based on the 2 years of data. These results also indicate that Eucalyptus plantations in China may be an important carbon sink due to the large plantation area.

Given the rapid increase in atmospheric CO₂ concentrations, estimation of the terrestrial carbon cycle in various biomes is urgently needed1,2. Because large amounts of carbon are retained in living plant and soil organic matter, the emission of these carbons into the atmosphere as CO₂ would have a profound effect on the global climate3. Recent interest in understanding how CO₂ emission from soils (or soil respiration) influences the global carbon cycle and its potential feedback on climate change has resulted in a growing number of studies4–6. Because soil respiration is the second largest carbon flux between terrestrial ecosystems and the atmosphere, releasing 10 times as much CO₂ to the atmosphere as the combustion of fossil fuels, it is imperative to investigate the impacts of human disturbance on ecosystem C dynamics, including soil respiration7,8.

Forests cover approximately 4.1 billion hectares of the Earth’s land surface and have been estimated to account for 80% of all aboveground carbon and approximately 40% of all belowground terrestrial carbon8, which suggests the importance of forests in mitigating additional atmospheric CO₂ emission9. In China, the total forest area is approximately 195 million hectares, and more than one-third of these forests are plantations10. Reforested plantations can sequester large amounts of CO₂ and offset the negative effect of fossil carbon emission11,12. When assessing ecosystems as carbon sinks or sources, a central concept is net ecosystem production (NEP), which is defined as the net annual carbon accumulation11,14. Although numerous studies have shown that NEP is affected by temperature, moisture, and stand age15,16, few studies have addressed how forest NEP is affected by different plant components, such as trees vs. understory plants.

Previous studies have indicated that understory vegetation, which is influenced by both resource quantity and resource heterogeneity17, greatly affects the properties and processes in forest ecosystems18,19. However, previous studies examining carbon cycling in forest ecosystems usually focused on the dominant overstory plant species.
Cumulative soil respiration. The cumulative soil respiration from March 2008 to March 2009 was highest in the control subplots (856 ± 116 C m⁻² yr⁻¹ in the 2-year-old plantations and 670 ± 68 g C m⁻² yr⁻¹ in the 24-year-old plantations) (Figure 1). The cumulative soil respiration in the 24-year-old plantations was lowest in the subplots with understory removal plus girdling (380 ± 86 C m⁻² yr⁻¹). The cumulative soil respiration in the 24-year-old plantations was lowest in the understory removal subplots (455 ± 57 g C m⁻² yr⁻¹) (Figure 1). RM ANOVA showed that understory removal plus girdling in the 2-year-old plantations and understory removal in the 24-year-old plantations significantly reduced the cumulative soil respiration compared with control plots (P = 0.02).

Fine root biomass. For both plantations, the fine root biomass was significantly reduced by girdling and by understory removal plus girdling (Figure 2). The fine root biomass was lowest in understory removal plus girdling plots, with values of 10.2 and 2.3 g m⁻² in the young and mature plantations, respectively. Understory removal did not significantly affect the fine root biomass in the 2-year-old plantations but significantly decreased the fine root biomass in the 24-year-old plantations (P = 0.018; Figure 2).

Ecosystem carbon storage. From 2008 to 2009, increases in the diameter at breast height (DBH), coarse root biomass, and understory biomass were greater for the 2-year-old plantations than for the 24-year-old plantations (Table 2). The net primary production (NPP) was greater for the 2-year-old plantations (1717.4 ± 123.11 g C m⁻² yr⁻¹) than for the 24-year-old plantations (924.24 ± 62.48 g C m⁻² yr⁻¹). The increase in understory plant biomass contributed 33.9% ± 0.08 to the NPP in the 2-year-old plantations and 14.1% ± 0.02 to the NPP in the 24-year-old plantations. The NPP values were consistent with the NPP values and were 1337.30 ± 54.89 g C m⁻² yr⁻¹ in the 2-year-old and 2009 (P = 0.04) and by 16% across both 2008 and 2009 (P = 0.07). Based on RM ANOVA, girdling did not affect soil respiration in the 24-year-old plantations but did affect soil respiration in the 2-year-old plantations (Table 1, Figure S1 and S2). In the 2-year-old plantations, girdling decreased soil respiration by 27% in 2008 (P = 0.02) and by 20% across 2008 and 2009 (P = 0.04).

Table 1: Pvalues for repeated measures ANOVAs concerning the effects of sampling time (T), understory removal (UR), girdling (G), and their interactions on soil respiration, soil temperature, and soil moisture content in 2008, 2009, and across 2008 and 2009 (designated as “Both”) in the 2-year-old and 24-year-old plantations

| Independent variable | Soil respiration | Soil temperature | Soil moisture content |
|-----------------------|------------------|------------------|----------------------|
|                       | 2008 | 2009 | Both | 2008 | 2009 | Both | 2008 | 2009 | Both |
| **2-year-old plantations** |      |      |      |      |      |      |      |      |      |
| T                     | <0.01 | <0.01 | <0.01 | <0.01 | <0.01 | <0.01 | <0.01 | <0.01 | <0.01 |
| T × UR                | 0.81 | 0.14 | 0.60 | 0.009 | 0.001 | 0.01 | 0.92 | 0.1 | 0.18 |
| T × G                 | 0.19 | 0.45 | 0.09 | 0.96 | 1 | 1 | 0.67 | 0.69 | 0.78 |
| T × UR × G            | 0.99 | 0.50 | 0.99 | 0.99 | 0.99 | 1 | 0.85 | 0.23 | 0.40 |
| UR                    | 0.13 | 0.04 | 0.07 | 0.17 | 0.003 | 0.02 | 0.67 | 0.61 | 0.90 |
| G                     | 0.02 | 0.12 | 0.04 | 0.94 | 0.57 | 0.72 | 0.72 | 0.87 | 0.81 |
| UR × G                | 0.64 | 0.49 | 0.56 | 0.60 | 0.46 | 0.48 | 0.59 | 0.25 | 0.47 |
| **24-year-old plantations** |      |      |      |      |      |      |      |      |      |
| T                     | <0.01 | <0.01 | <0.01 | <0.01 | <0.01 | <0.01 | <0.01 | <0.01 | <0.01 |
| T × UR                | 0.01 | <0.01 | <0.01 | 0.002 | <0.01 | <0.01 | 0.09 | 0.82 | 0.41 |
| T × G                 | 0.30 | 0.19 | 0.37 | 0.89 | 0.98 | 0.99 | 0.90 | 0.55 | 0.62 |
| T × UR × G            | 0.17 | 0.52 | 0.43 | 0.97 | 0.93 | 0.99 | 0.96 | 0.77 | 0.91 |
| UR                    | 0.08 | 0.001 | 0.01 | 0.80 | 0.27 | 0.54 | 0.56 | 0.37 | 0.51 |
| G                     | 0.48 | 0.11 | 0.29 | 0.84 | 0.67 | 0.80 | 0.41 | 0.95 | 0.55 |
| UR × G                | 0.15 | 0.12 | 0.14 | 0.60 | 0.61 | 0.59 | 0.98 | 0.68 | 0.80 |
the second year 31,32. The different responses to girdling in girdling reduced respiration by 56% in the first year and 65% in caused even greater reductions in soil respiration; that is, tree resprouting trait of Eucalyptus, which allows roots to live and Eucalyptus plantations vs. boreal forests can be attributed to the

plantations and 420.15 ± 120.75 g C m⁻² yr⁻¹ in the 24-year-old plantations (Table 2). The plant biomass was greater in the 24-year-old plantations than in the 2-year-old plantations (P < 0.01), but soil organic carbon and floor litter carbon did not differ between the young and old plantations (Figure 3). The total carbon storage also did not differ between the 2-year-old plantations and the 24-year-old plantations (P = 0.16, Figure 3d), with values of 21.2 ± 0.98 and 24.4 ± 1.51 kg C m⁻² yr⁻¹, respectively. Compared with other dominant forest types in China, the NPP and NEP of Eucalyptus forests were significantly higher than the other forest types and the average values for China’s forest ecosystems (Table 3).

Discussion

Responses of soil respiration to tree girdling and understory removal. In the first year after tree girdling, soil respiration was decreased by 27% in the 2-year-old plantations and by 22% in the 24-year-old plantations. Decreases in respiration in the first year after girdling were greater in the current study than in previous studies in Eucalyptus plantations, in which respiration was reduced by only 14%-29. The girdling of trees in boreal forests, however, caused even greater reductions in soil respiration; that is, tree girdling reduced respiration by 56% in the first year and 65% in the second year31-32. The different responses to girdling in Eucalyptus plantations vs. boreal forests can be attributed to the resprouting trait of Eucalyptus, which allows roots to live and continue to respire for more than 1 year after girdling. Our previous study showed that that 51% and 62% of the fine roots remained alive 1 year after girdling in the 2-year-old and 24-year-old plantations, respectively32; significant reductions in fine root biomass were not detected until 32 months after girdling. The dead roots also represent a new carbon input to soil and are important substrates supporting the heterotrophic respiration of soil microorganisms35.

Understory removal in the first year reduced soil respiration by 14% and 34% compared to the control in the 2-year-old plantations and the 24-year-old plantations, respectively, suggesting that understory plant roots made substantial contributions to soil respiration36. In our previous study, understory plant removal in a plantation of mixed native tree species decreased annual soil respiration by only 6%34. We hypothesize that the effect of understory removal on soil respiration depended on species composition, plantation age, and stand structure35. For example, soil respiration was 4.2% higher in an oak forest with understory than without understory, whereas soil respiration was 22.6% higher in a scot pine forest with understory than without understory36.

Research has suggested that understory removal should increase soil water content due to decreased transpiration and that soil temperature should increase due to increased light penetration37. In our experiment, however, understory removal and tree girdling did not affect soil water content. This might be explained by a possible trade-off between water use by trees and understory plants; a previous study reported that introduced Eucalyptus can consume more water than native species38. The soil temperature increased only during the second year of our study and only after understory removal in the 24-year-old plantations. Because the treatments only marginally affected soil temperature and moisture, we inferred that soil temperature and soil moisture were not major factors affecting soil respiration in the present study. Our results indicated that nutrient availability partially explained the dynamics of soil respiration because the soil NO₃⁻-N significantly increased after understory removal (Figure S4). We postulated that understory removal eliminated the nutrient translocation from soil to understory plants, consequently reducing the total soil respiration32.

The contribution of understory plants to ecosystem carbon storage. Understory species such as Dicranopteris often form a dense mat under the open canopy in tropical regions39. In both plantations in our study, however, the canopies are open, and
understory plants receive abundant solar energy and grow rapidly. A recent study also showed that light heterogeneity directly influences the understory plant community. Our results confirm that increases in plant biomass would be underestimated if understory plants are ignored. The biomass of understory plants accounted for 20% and 31% of the total plant biomass in the young and mature plantations, respectively. These values would clearly be lower in a natural forest with a closed canopy. However, in subtropical China, the understory layer could account for 10–19.9% of the total forest biomass and should not be ignored when assessing total carbon pools in the forest ecosystems of subtropical China.

Effects of stand age on ecosystem carbon storage. The NEP was reported to be negative in young plantations because, unlike mature plantations, carbon emission can exceed carbon storage in younger plantations. Photosynthesis-derived carbon is usually low in the early stages of reforested plantations, which have a small leaf area index, and it then increases over time. For instance, gross ecosystem photosynthesis across a boreal jack pine chronosequence was only

| Variable                              | 2-year-old plantations | 24-year-old plantations |
|---------------------------------------|------------------------|-------------------------|
| ΔDBH (cm)                             | 2.35 (0.10)a           | 0.56 (0.08)b            |
| ΔHeight (m)                           | 2.33 (0.52)a           | 1.44 (0.26)a            |
| ΔBab [g DW m⁻² yr⁻¹]                  | 43.1 (8.46)a           | 619.2 (51.9)a           |
| ΔBc [g DW m⁻² yr⁻¹]                   | 1471.8 (296.8)a        | 832.1 (121.8)a          |
| ΔBu [g DW m⁻² yr⁻¹]                   | 352.9 (43.72)a         | 153.6 (21.8)b           |
| ΔBu [g DW m⁻² yr⁻¹]                   | 241.9 (51.4)a          | 129.3 (7)a              |
| NPP [g C m⁻² yr⁻¹]                    | 1260.4 (444.2)a        | 288.3 (56.6)b           |
| Rh [g C m⁻² yr⁻¹]                     | 380.08 (86.27)a        | 504.09 (62.15)b         |
| NEP [g C m⁻² yr⁻¹]                    | 1337.30 (54.89)a       | 420.15 (120.75)b        |

Figure 3 | Soil organic carbon (a), plant biomass carbon (b), litter carbon (c), and total carbon (d) in the 2-year-old and 24-year-old plantations. Values shown are the mean ± 1 SE; n = 3. Within each panel, means with different letters are significantly different (P < 0.05).
left on the stand and burned. Next, the Eucalyptus plantations were established. In these plantations, Eucalyptus saplings were planted with a spacing of 3 m × 2 m and a density of approximately 1660 trees ha⁻¹. Six experimental Eucalyptus plantations on homogenous degraded hilly areas were used. Three plantations were 24 years old (established in 1984), and the remaining plantations were 2 years old (established in 2006). Understory species in both plantations are dominated by Dicranopteris dichotoma, other understory species include Rhododendron tomentosum, Baeckea frutescens, Dianella ensifolia, Wkstroemia indica, and Blechnum orientale. Indigenous tree species are rare within the study site because of seed resource limitation and dense understory plants. In March 2009, the mean understory biomass was 772 ± 92 g dry weight m⁻² in 2-year-old plantations and 2116 ± 61 g dry weight m⁻² in 24-year-old plantations. The vegetation was considered to consist of two functional groups: overstory Eucalyptus trees and the Dicranopteris dichotoma-dominated understory.

**Experimental design.** One experimental plot (10 m × 10 m) was established in each plantation in December 2008. Each plot was divided into four subplots; each of the four subplots corresponded to one treatment. The randomized block design had two levels for each of the two factors (±girdling and ±understory removal). There were four treatment combinations: no girdling and no understory removal (CK, or control); girdling but no understory removal (G); no girdling but understory removal (UR); and girdling plus understory removal (GUR). A 40-cm-deep trench was created around each subplot to eliminate the intrusion of roots from the other subplots. Understory removal and tree girdling were performed in March 2008. Understory plants were manually removed with a machete. For girdling, we cut 10-cm bands around the stem of each Eucalyptus tree in the designated plots at 50 cm above the soil surface. Eucalyptus trees can resprout after girdling and Dicranopteris can grow from remnant roots, the new growth of Eucalyptus and Dicranopteris (and growth of any other understory plant) was removed monthly.

**Soil CO₂ efflux.** Soil respiration was measured monthly from March 2008 to December 2008 and bimonthly from January 2009 to March 2010 between 9:00 a.m. and 12:00 a.m. with an LI-8100 automated soil CO₂ flux system (LI-COR Inc., Lincoln, NE, USA). To measure soil respiration, three PVC collars (20 cm diameter and 5 cm height) were placed at a depth of 2 cm in each subplot, and small living plants in the soil collars were removed by hand. PVC collars remained fixed throughout the experiment. Soil temperature at 5-cm depth and volumetric soil moisture were measured by probes attached to the automated CO₂ measurement device when the respiration was recorded. The soil respiration in plots with understory removal plus tree girding was considered to be heterotrophic respiration (Rₑ). Cumulative soil respiration for each treatment was determined by summing the amount of soil respiration with the number of days between sampling times.

**Net primary production and net ecosystem production.** Biometric approaches were used to estimate net primary production (NPP, g C m⁻² yr⁻¹) and net ecosystem production (NEP, g C m⁻² yr⁻¹) (for details, see Chen et al. 2011). Briefly, seven typical trees were harvested to calculate total dry biomass. Allometric relationships between tree biomass and the diameter at breast height (DBH, 1.3 m) and tree height were determined. The DBH and height of all trees within the plots were measured in April 2008 and April 2009. At the same time, to estimate understory plant biomass, we harvested the aboveground and belowground parts of all understory plants in two 1 × 1 m subplots in each of the three replicated plots of each plantation. Understory plants were dried to constant mass at 75°C. The value for carbon concentration was 0.4577.

Annual litter input was determined by deploying three nylon-mesh litter traps (1 m × 1 m) at each plantation from July 2008 to July 2009. Litter was collected monthly and dried to a constant mass at 75°C. The biomass of fine roots (diameter < 2 mm) was determined by collecting five soil cores every 3 months (5 cm diameter, 20 cm depth) in the subplots from March 2009 to March 2010 and a final time in November 2010 to investigate the effects of girdling and understory removal on fine root biomass. In the laboratory, fine roots were removed from the soil, rinsed with deionized water, and dried to constant mass. The sum of biomass increases in aboveground parts, roots, understory plants, and litterfall was considered to be NPP. The cumulative soil respiration in plots with understory removal plus tree girding was considered to be cumulative heterotrophic respiration (Rₑ). The difference between NPP and Rₑ was considered to be NEP.

The sum of soil organic carbon, floor litter carbon, and stand plant biomass was considered ecosystem carbon storage. For the determination of floor litter carbon storage, two 1 m × 1 m subplots were selected in each of the plots of the plantation in 2008. The floor litter was collected and dried to a constant mass at 75°C. Data for soil organic carbon and soil bulk density were obtained from Wu et al. (2011b), who conducted research in the same plots.

**Statistical analyses.** We used one-way analysis of variance (ANOVA) to analyze the effects of treatments on soil respiration, soil temperature, soil moisture, ecosystem carbon storage, NPP, and NEP. Repeated measures analysis of variance (RM ANOVA) was used to determine the effects of sampling time, understory removal, and girdling on soil respiration, soil temperature, and soil moisture content. The Bonferroni post hoc test was used to compare the means of different treatments. The carbon pools were considered to be time- and site-dependent, and the within-subject effects were sampling time and its interaction with understory removal and girdling. The statistical analyses were conducted using SPSS 15 (SPSS, Inc., Chicago, IL, USA). Differences were considered significant at the 0.05 level.
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Author contributions

J.W., Z.F.L. and S.L.F. conceived and designed the experiments. J.P.W., G.M.H., D.M.C. and S.Z.W. performed the experiments. J.P.W., W.X.Z., D.M.C. and Y.H.S. analyzed the data. J.P.W., S.L.F. and Z.F.L. wrote the manuscript.

Additional information

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