Developmental exposures to bisphenol S, a BPA replacement, alter estrogen-responsiveness of the female reproductive tract: A pilot study

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Abstract: Developmental exposures to bisphenol A (BPA), an estrogen receptor agonist, can disrupt development of the female reproductive tract in rodents and non-human primates. Due to an increased public knowledge of negative health effects associated with BPA exposure, BPA has begun to be phased out of many consumer products and in some cases it has been replaced with structurally similar compounds including bisphenol S (BPS). This study examined CD-1 mice exposed to a low dose of BPS during early development (200 μg/kg/day from gestational day 8 until postnatal day 19). BPS altered expression of estrogen-responsive genes in both the uterus and ovary, and induced increases in ovarian follicular development in pre-pubertal females evaluated at postnatal day 22. Prior studies have revealed that developmental exposures to environmental chemicals including BPA alter the response of animals to hormonal or carcinogen challenges experienced later in life. To evaluate whether early life exposures to BPS alter responses of females to an estrogen challenge, additional females were exposed to ethinyl estradiol from...
postnatal day 19 through postnatal day 21. BPS-treated females responded abnormally to this estrogen challenge, displaying heightened responses in the uterus and diminished responses in the ovary. Although additional studies are needed to characterize the mechanisms by which BPS alters the female reproductive tract, this pilot study provides evidence that a common BPA replacement chemical may have endocrine disrupting properties.

**Subjects:** Environmental & Ecological Toxicology; Reproduction; Developmental Biology; Endocrinology

**Keywords:** apoptosis; endocrine disruptor; estrogen receptor; ethinyl estradiol; ovarian follicles; puberty; proliferation; uterine endometrium

1. **Introduction**

It is widely understood that early life exposures to synthetic estrogens can induce alterations to the female reproductive tract, some of which will manifest at puberty (Crain et al., 2008; Mandrup et al., 2013; Soto, Rubin, & Sonnenschein, 2007). One of the best known examples of this comes from human exposures to the pharmaceutical diethylstilbestrol (DES), which was prescribed to pregnant women to prevent spontaneous abortion (McLachlan, 2006). Girls exposed to DES during gestation, so-called DES daughters, often developed malformations of the reproductive tract and had an increased risk of developing clear cell adenocarcinoma of the vagina starting at puberty (Herbst, Ulfelder, & Poskanzer, 1971; Hoover et al., 2011). Mice exposed to DES during early development are similarly affected, developing lesions of the reproductive tract that become obvious at puberty and in adulthood (Newbold, Bullock, & McLachlan, 1990; Newbold, Jefferson, & Padilla-Banks, 2007; Suzuki et al., 2002). Rodent models used in DES studies proved to be highly predictive of other adverse outcomes observed in DES daughters such as breast cancer (Soto, Vandenberg, Maffini, & Sonnenschein, 2008).

A relatively large number of estrogenic endocrine disrupting chemicals (EDCs) have now been identified. Both controlled laboratory animal and epidemiological data suggest adverse effects on the female reproductive system as a result of early life exposures (Bergman et al., 2013; Gore et al., 2015; Hunt, Susiarjo, Rubio, & Hassold, 2009; Maffini, Rubin, Sonnenschein, & Soto, 2006; Rochester, 2013). In a recent review, Peretz and colleagues assessed all studies published from 2007 to 2013 that examined the impact of one xenoestrogen, bisphenol A (BPA), on the male and female reproductive systems (Peretz et al., 2014). These authors concluded that BPA induced or was associated with disturbances of the ovary, oviduct and uterus in laboratory animals, and in human epidemiology studies. Another systematic review of all published studies reporting uterine effects in laboratory animals exposed to BPA during the period of uterine organogenesis and differentiation concluded that there is clear evidence for effects of early, developmental exposure to BPA leading to uterine abnormalities during adulthood in mice and rats (Suvorov & Waxman, 2015). Further, consistent effects of BPA on a number of reproductive endpoints have been shown across studies conducted in multiple laboratories, and in different species (Vandenberg, Ehrlich et al., 2013).

Bisphenol S (BPS) is a chemical that is structurally similar to BPA (Rosenmai et al., 2014). Because both scientific and public knowledge of adverse outcomes related to BPA exposure have increased (Chapin et al., 2008; Vandenberg, Hunt, Myers, & vom Saal, 2013), consumer demand for BPA-free products has encouraged the replacement of BPA with other related compounds including BPS (Mathew, Sreedhanya, Manoj, Aravindakumar, & Aravind, 2014; Rochester & Bolden, 2015; Vandenberg, Luthi, & Quinerly, 2017). BPS is now used in a range of consumer products including thermal receipts, food-contact paper products, and canned foods (Liao & Kannan, 2013; Liao, Liu, & Kannan, 2012; Liao, Liu, Alomirah et al., 2012; Liao, Liu, Guo et al., 2012; Liao, Liu, Moon et al., 2012; Mathew et al., 2014). Biomonitoring studies reveal that human exposures to BPS are likely to be widespread (Ye et al., 2015); approximately 97% of individuals in the US have detectable levels of...
BPS metabolites in their urine (Liao, Liu, Alomirah et al., 2012). Back-calculations from these urinary concentrations suggest daily exposures in the range of 0.3–2 μg/day, however, these exposures will likely rise as the replacement of BPA in various consumer goods also increases. These back-calculations are based on limited toxicokinetic and exposure data and assumptions about clearance, and may therefore be an underestimate (Lv et al., 2017). Like BPA, human exposure to BPS appears to occur predominantly by two routes of exposure: absorption through the skin (Porras, Heinälä, & Santonen, 2014) and ingestion from plastic leachates, can linings, and dust (Bittner, Yang, & Stoner, 2014; Liao, Liu, Guo, et al., 2012).

In vitro studies show that BPS acts as an agonist for estrogen receptor (ER) α, ERβ, and membrane ER (Grignard, Lapenna, & Bremer, 2012; Kuruto-Niwa, Nozawa, Miyakoshi, Shiozawa, & Terao, 2005; Molina-Molina et al., 2013; Vinas & Watson, 2013). Studies in zebrafish support the hypothesis that this chemical can disrupt both reproductive and developmental endpoints through endocrine pathways (Ji, Hong, Kho, & Choi, 2013; Kinch, Ibhaeziebo, Jeong, Habibi, & Kurrasch, 2015; Naderi, Wong, & Gholami, 2014; Qiu et al., 2016). Studies observing effects of BPS in mammals are limited, although there is evidence from our laboratory and others that it can disrupt maternal behaviors and induce uterotrophic responses, consistent with its suspected estrogenic properties (Catanese, Suvorov, & Vandenberg, 2015; Catanese & Vandenberg, 2016; Yamasaki, Noda, Imatanaka, & Yakabe, 2004). There is also evidence that BPS exposure can affect body weight and neurobehaviors in developmentally exposed male offspring (Kim, Colon, Chawla, Vandenberg, & Suvorov, 2015).

Based on previous studies which observed changes in estrogen dependent endpoints due to early life BPA or BPS exposure, we hypothesized that developmental exposure to BPS would induce changes in estrogen dependent endpoints specifically of the female reproductive tract, with effects observed on a cellular and molecular level. We conducted a pilot study examining mice exposed to 200 μg BPS/kg/day from gestational day 8—postnatal day [PND]19. This dose was selected because it is 1/50th of the toxicological NOAEL of 10 mg/kg/day (EPA U, 2014). Developmental exposures to BPS resulted in modest alterations to gene expression in the female reproductive organs in the pre-pubertal mouse. Because many of the effects of DES on the female reproductive tract were not obvious until puberty, we also provided pre-pubertal females (treated with and without BPS during early development) with an estrogen challenge that was sufficient to induce an uterotrophic response. We noted that females perinatally exposed to BPS (referred to as BPS + EE females) responded to this estrogen challenge differently from controls (referred to as control + EE females), suggesting that BPS alters the responsiveness of the female reproductive tract to an estrogen challenge. These observations may be similar to how a female developmentally exposed to BPS will respond to natural pubertal hormones, and indicate that widespread exposures to BPS may have implications for public health.

2. Results

2.1. Effects of perinatal BPS exposures on the uterus and oviduct at PND22

2.1.1. Tissue morphology

A number of characteristics of tissue organization in tissues of the female reproductive tract are known to be influenced by estrogen exposure including height of the uterine epithelium, e.g. the endometrium. To characterize the effects of developmental BPS exposure on the uterus, the height of the uterine epithelium was measured at multiple distinct locations along the endometrial layer. Perinatal exposure to BPS had no effect on tissue organization of the uterus including endometrial cell height (Figure 1(A) and (B)).

Tissue organization of the oviduct was analyzed similarly, although both cell height and width were measured. In the oviduct, no significant qualitative or quantitative changes were observed for epithelial height, epithelial width, or height:width ratios of the oviduct epithelium (Figure 1(C) and (D) and data not shown).
2.1.2. Apoptosis and proliferation

To determine whether developmental BPS exposures alter other cellular characteristics in the uterus, the number of cells expressing Ki67, a marker of proliferation, was quantified in both the epithelial and stromal compartments (the endometrium and lamina propria, respectively). No significant effects of BPS were noted in either tissue compartment (Figure 2(A) and (B)).

Similarly, apoptotic cells were quantified using positive TUNEL (Terminal deoxynucleotidyl transferase (TdT) dUTP Nick-End Labeling) staining. No significant effects of BPS were noted (Figure 2(A) and (B)).
Ki67 and TUNEL expression were also analyzed in oviduct epithelial tissue. No effects of BPS treatment were observed for the number of epithelial cells in the oviduct expressing either the proliferation or apoptotic markers (Figure 2(C)).

2.1.3. Gene expression
To quantify changes in uterine development influenced by early life BPS exposure, several genes involved in hormone responsiveness and known to be sensitive to estrogens were measured with the use of qPCR. These genes included: ESR-1 (estrogen receptor α), ESR-2 (estrogen receptor β), IGF-1 (insulin growth factor 1) and EGF-R (epidermal growth factor 1). Perinatal BPS exposure significantly increased the expression of IGF-1 (p < 0.05) (Figure 3) but not EGF-R in the uterus. A non-significant decrease in ESR-1 was also observed (Figure 3). Expression levels of ESR-2 were very low, precluding accurate quantification (data not shown). Gene expression in the oviduct was not assessed due to a lack of available qPCR grade oviduct tissue.
2.2. Altered response in the uterus of BPS-exposed females to an estrogen challenge

Prior studies have revealed that developmental exposures to EDCs, including BPA, alter the response of animals to hormonal or carcinogen challenges experienced later in life (Betancourt, Eltoum, Desmond, Russo, & Lamartiniere, 2010; Durando et al., 2007; Prins, Birch, Tang, & Ho, 2007; Wadia et al., 2007). Here, one female from each litter was treated with 1 μg/kg/day ethinyl estradiol (EE) from PND19-PND21, prior to the onset of puberty (Supplemental Figure 1). This dosing paradigm is sufficient to induce an uterotrophic response in prepubertal mice, and thus can artificially induce some physiological responses consistent with puberty (Vandenberg et al., 2006).

2.2.1. Tissue morphology

In control + EE females, the EE challenge induced slight increases in uterine epithelial height, but these increases were not significantly different from unchallenged control animals (Figure 4(A)). Although the endometrium is known to be responsive to estrogen, prior studies suggest that higher doses are needed to induce statistically significant increases in cell height (Vandenberg et al., 2006). Yet, surprisingly, in BPS + EE females, a significant increase in uterine epithelial height was observed.
following the pubertal estrogen challenge (Figure 4(A)). This result suggests that perinatal BPS exposure may act to enhance the uterine response to EE at puberty.

2.2.2. Apoptosis and proliferation

We again evaluated proliferation and apoptosis using immunohistochemistry for Ki67 and TUNEL labeling, respectively. In both control + EE and BPS + EE females, the EE challenge did not significantly affect the number of cells proliferating or undergoing apoptosis in the uterine stroma or epithelium, regardless of whether the animals were perinatally exposed to BPS (Figure 4(B)).

2.2.3. Gene expression

Finally, expression of IGF-1, EGF-R and ESR1 were evaluated in uterine samples following an EE challenge. In control + EE females, treatment with EE had no significant effects on uterine gene expression (Figure 4(C)), although a relative decrease of ESR-1 expression was observed. No effects on uterine gene expression were seen in BPS + EE females.

2.3. Effects of perinatal BPS exposures on the ovary at PND22

2.3.1. Tissue morphology

We next characterized the morphology of the ovary by counting the number of primordial, primary, secondary, and antral follicles in sections collected from animals exposed to BPS and unexposed controls (Figure 5(A)). Follicles were only counted if the nucleus of the oocyte was visible. Perinatal exposure to BPS significantly increased the number of secondary follicles compared to unexposed controls (Figure 5(B)). No significant effects were observed on other follicle stages or total follicles (Figure 5(B) and (C)).

Although there were no statistically significant differences in the number of sections evaluated between treatment groups, the total number of sections (and thus the size of the ovaries) was variable between individuals (Table 1). To account for possible differences in ovarian size, ovarian follicles were counted in four sections separated by at least 50 μm. The BPS-treated animals displayed modest increases in the number of primary, secondary and antral follicles (Figure 5(D)) and total...
follicles (Figure 5(E)), but only the increase in secondary follicles was statistically significant. Atretic follicles were not observed in any animal in any treatment group (data not shown).

2.3.2. Apoptosis and proliferation
Ovarian tissue was assessed to characterize the effects of BPS on proliferation and apoptosis. BPS exposure was associated with modest increased expression of Ki67 in both granulosa cells and theca cells, although neither of these differences achieved statistical significance (Figure 6(A) and (B)). No effect of BPS was seen on TUNEL expression.

Table 1. Average number of sections containing ovary

| Treatment group (perinatal/EE) | # Sections (n ± SEM) | Range (min, max) |
|-------------------------------|---------------------|------------------|
| Control + no challenge        | 117.0 ± 12.6        | 89, 143          |
| BPS + no challenge            | 115.4 ± 13.2        | 89, 151          |
| Control + EE                  | 125.0 ± 11.9        | 110, 160         |
| BPS + EE                      | 92.4 ± 18.0         | 59, 140          |
Figure 6. No significant effects of BPS on expression of markers of proliferation or apoptosis in the ovary. (A) Expression of Ki67 and TUNEL labeling in the ovary. Granulosa cells are indicated by an arrow and theca cells are indicated by an arrowhead. (B) Quantification of expression revealed non-significant increases in Ki67 expression in both cell types in BPS-exposed ovaries. No differences were observed for TUNEL expression in BPS-treated ovaries. Photomicrographs were collected with a 20× EpiPlan Objective. (C) Expression of four genes in the ovary at PND22.

Notes: In all panels, * indicates \( p < 0.05 \), independent samples t-test. In all groups, \( n = 5 \) per treatment.
2.3.3. Gene expression

To quantify changes in ovarian development influenced by early life BPS exposure, the same genes that are known to be crucial for hormone regulation in the uterus were measured with the use of qPCR. Perinatal BPS exposure significantly increased the expression of IGF-1 (\(p < 0.05\)) and EGF-R (\(p < 0.05\)) in the ovary (Figure 6(C)). An increase in the expression of ESR-1 and ESR-2 was also observed in BPS-treated females, although this was not statistically significant compared to controls (Figure 6(C)). Overall, BPS induced increases in the expression of genes which are drivers of sexual maturity and regulate estrogen dependent pathways.

2.4. Altered response in the ovary of BPS-exposed females to an estrogen challenge

2.4.1. Tissue morphology

In control + EE females, the pubertal EE challenge significantly increased the number of primary, secondary and antral follicles, as well as the total number of follicles, when measured in the entire ovary (Figure 7(A) and (B)). When only four sections were evaluated in control + EE females, the EE challenge significantly increased the number of primary, secondary and antral follicles (Figure 7(C)) and total follicles (Figure 7(D)). This result is consistent with an estrogen-induced ovarian response, similar to what is seen in the maturation of follicles at the onset of puberty. In contrast, BPS + EE
females saw no EE-induced increases in the number of total follicles, nor were any effects on specific follicle categories observed (Figure 7(A) and (B)). This was also true when only four sections were evaluated (Figure 7(C) and (D)). These results suggest that an estrogen challenge is ineffective at inducing an ovarian response in BPS-exposed females.

2.4.2. Proliferation and apoptosis

No changes in cell proliferation or apoptosis markers were observed in either the granulosa cells or the theca cells in the ovary of control + EE females following an estrogen challenge (Figure 8(A) and (B)). In contrast, BPS + EE females had increased numbers of theca cells that were positive for Ki67 and TUNEL, although only the apoptotic marker was statistically significant (Figure 8(A) and (B)).

Figure 8. Significant alterations in apoptosis and gene expression in the ovary of BPS-treated females after an estrogen challenge. (A) Quantification of Ki67 in the ovary of control + EE and BPS + EE females after an estrogen challenge. (B) Quantification of TUNEL in the ovary of control + EE and BPS + EE females after an estrogen challenge. (C) An estrogen challenge decreased the expression of estrogen-sensitive genes in the ovaries of BPS + EE females.

Notes: In all panels, * indicates \( p < 0.05 \), independent samples \( t \)-test comparing EE-challenged with non-challenged females from the same perinatal treatment group (control or BPS). \( \delta \) indicates significant differences, \( p < 0.05 \), between control + EE and BPS + EE females. In all groups, \( n = 5 \) per treatment.
2.4.3. Gene expression

Control + EE females showed a relative (but non-significant) increase in ESR-1 expression in the ovary when compared to animals not provided with an EE challenge, suggesting that an estrogen challenge increases the expression of ERα in this organ (Figure 8(C)). In contrast, BPS + EE females experienced significant decreases in the expression of EGF-R and ESR-2 compared to BPS-treated females that were not given an estrogen challenge (Figure 8(C)).

3. Discussion

Here, we have shown that perinatal exposures to BPS, a common BPA replacement, can disrupt development of the mouse female reproductive tract. Relatively subtle effects were observed at PND22 in animals exposed to BPS alone, with statistically significant effects in the uterus limited to gene expression and effects in the ovary seen in both tissue morphology (e.g. number of secondary follicles) and gene expression. Further, when animals were provided with three days of estrogen treatment at a dose sufficient to induce a uterotrophic response, early life BPS exposure was associated with altered gene expression and abnormal tissue morphology in both the uterus and ovary.

In the pre-pubertal female mouse, the most striking effects of BPS were observed on gene expression endpoints, with significant effects on the expression of IGF-1 in the uterus and ovary and EGF-R in the ovary (Figures 3 and 6). IGF-1 was selected for analysis because prior studies have shown that estrogens can regulate expression of IGF-1 in uterine stromal cells and ovarian granulosa cells (Ogo et al., 2014). EGF-R expression is also critically regulated by estrogen and expression of this gene increases late in folliculogenesis, consistent with our finding of increased numbers of mature follicles in BPS treated animals (El-Hayek, Demeestere, & Clarke, 2014; Garnett, Wang, & Roy, 2002). Our results suggest that BPS may be able to act as an estrogen to stimulate increased expression of IGF-1 and EGF-R. Although the effects of BPS on ESR1 and ESR2 expression (encoding ERα and ERβ, respectively) were not statistically significant, likely due to the small sample size in this pilot study, their increased expression in BPS-treated ovaries is also consistent with BPS acting as an estrogen. One caveat that must be considered in the analysis of gene expression data is that these are evaluations of whole-organ cell lysates; because different gene expression is anticipated with different follicular stages, the changes in gene expression may be reflective of changes in follicular stage, including changes that we could not detect with our histological method (see further discussion below).

It has been reported previously that developmental exposure to estrogenic EDCs such as BPA and DES result in abnormal ovarian follicle formation (Hunt et al., 2012; Peretz, Gupta, Singh, Hernández-Ochoa, & Flaws, 2011; Rivera, Varayoud, Rodriguez, Muñoz-de-Toro, & Luque, 2011; Rodriguez, Santambrosio, Santamaria, Muñoz-de-Toro, & Luque, 2010) and tissue organization in the uterus (Markey, Michaelson, Veson, Sonnenschein, & Soto, 2001; Markey, Wadia, Rubin, Sonnenschein, & Soto, 2005; Newbold, 2004; Newbold, Jefferson, & Padilla-Banks, 2009; Newbold et al., 2007). Here, we measured significant increases in the number of secondary follicles in ovaries of BPS-treated females (Figure 5), although no significant effects were observed on tissue organization, tissue morphology, cell proliferation or apoptosis in the uterus or oviduct (Figures 1 and 2). This may be due to the limited size of this pilot study (n = 5 per group) or the selection of a single BPS dose (200 μg/kg/day). Alternatively, these results could provide evidence that BPS induces alterations at the molecular level of biological organization in the uterus, but that more time is needed to observe effects at the level of the tissue. Additional studies are also needed to further explore the effects of BPS on the ovary, specifically to evaluate animals throughout adulthood. In particular, additional studies should examine not only the phenotype of the ovaries in BPS-exposed females (e.g. follicle number, follicle maturity, gross abnormalities etc.) they should also measure the quality and developmental competence of oocytes. If the number of mature follicles formed after puberty is significantly reduced in BPS-treated animals, this might indicate additional effects on measures of fecundity and fertility. Not only do these measures represent obvious adverse outcomes, they suggest possible endpoints with public health implications.
Importantly, a critical part of our experimental design included a pre-pubertal estrogen challenge. To our knowledge, no studies using developmental EDC exposures and a pre-pubertal estrogen challenge have been conducted using ovarian endpoints, making it difficult to make direct comparisons with our study. However, prior studies have revealed that developmental exposures to EDCs including BPA and DES alter the response of animals to hormonal or carcinogen challenges experienced later in life (Betancourt et al., 2010; Durando et al., 2007; Newbold, Jefferson, Padilla-Banks, & Haseman, 2004; Prins et al., 2007; Wadia et al., 2007). In our study, one female from each litter was treated with 1 μg/kg/day EE for three days, from PND19-PND21; this dosing paradigm was selected because it is sufficient to induce an uterotrophic response in prepubertal mice, and thus can artificially induce some physiological responses consistent with puberty (Vandenberg et al., 2006). We found that the estrogen challenge induced significant follicular maturation in control + EE females at PND22, with specific increases in the number of secondary and antral follicles, and the total number of follicles (Figure 7). These results are consistent with what is expected to occur in the ovary at the onset of puberty, when more advanced follicles are visible in the ovary (Monget et al., 2012; Omoto, Lathe, Warner, & Gustafsson, 2005). Although it is possible that EE prevents follicular loss, it is more likely that EE advances development of the ovary, increasing the number of mature follicles that are more likely to be counted with our method, which only evaluates sections every 50 μm and thus biases our counts toward larger follicles. Of course, these values are not counts of all follicles in the ovary; the method we use likely under-samples smaller follicles (primordial and primary), but the sampling methods are consistent between ovaries, allowing values to be compared between individuals. It is therefore not possible to determine whether BPS alters the complete pool of primordial follicles. This is potentially a sensitive endpoint that should be investigated in the future.

Strikingly, the increased number of mature follicles observed in control + EE females compared to controls that were not given an estrogen challenge were not observed in BPS + EE females; in these animals, the estrogen challenge did not appear to increase follicular maturation in the ovary. These results suggest that perinatal exposures to BPS may reduce responses of the ovary to estrogens at puberty (Figure 7), even when BPS exposure appears to heighten responses of the uterus when presented with an estrogen challenge (Figure 4(A)). Together, results observed in the ovary and uterus suggest that these organs respond in different ways to an estrogen challenge after perinatal BPS exposure. Additional analyses that examine additional aspects of the hypothalamic-pituitary-ovarian axis including gene expression in the hypothalamus and pituitary may help to elucidate the mechanisms by which BPS and similar compounds induce distinct alterations in estrogen-sensitive organs including the ovary and uterus.

BPA has a wide range of molecular behaviors including estrogen receptor agonist and antagonist behavior (both ERα and ERβ), an ability to bind to membrane ERα and the transmembrane estrogen receptor GPR30, actions as a thyroid hormone antagonist and an androgen receptor antagonist, and ability to bind to the estrogen related receptor (ERRy) and the aryl hydrocarbon receptor (AhR) (Vandenberg, Maffini, Sonnenschein, Rubin, & Soto, 2009). In contrast, most studies of BPS point to its estrogen receptor agonist properties (via ERα, ERβ and membrane ERs) (Grignard et al., 2012; Kuruto-Niwa et al., 2005; Molina-Molina et al., 2013; Vinas & Watson, 2013). Some of these studies suggest that BPS may be a more potent estrogen mimic than BPA, at least for some endpoints. Considering those molecular data, previous studies showing that BPS can disrupt a range of endpoints in zebrafish and rodents (Catanese & Vandenberg, 2016; Kim et al., 2015; Kinch et al., 2015; Naderi et al., 2014), and the data we present in this manuscript, it is clear there is a need for further studies to observe additional adverse outcomes associated with BPS exposure; it is equally clear that the data are not supportive of safety for this replacement chemical.

3.1. Conclusions
In the last year, a number of studies examining the effects of BPS have suggested that developmental exposures to this compound induce altered health outcomes through endocrine disrupting actions in rodents and zebrafish (Rochester & Bolden, 2015). The results we have acquired here are consistent with the hypothesis that developmental exposure to BPS disrupts cellular and molecular
events in the developing mammalian female reproductive tract. The effects on the ovary that we observed were most pronounced after an estrogen challenge; we found that BPS induces premature ovarian development, but diminishes the response of the ovary from being further stimulated by estrogens. In contrast, the effects of BPS on uterine tissue organization are only apparent after an estrogen challenge meant to mimic the onset of puberty. Further studies are needed to extend this pilot study, both to examine additional endpoints of interest, to include older ages, and to examine apical endpoints related to fertility, fecundity, and reproductive health.

4. Materials and methods

4.1. Animals

Adult outbred CD-1 mice of both sexes were obtained from Charles River Breeding Laboratories (Raleigh, NC). The CD-1 strain has previously been shown to be a sensitive model for the use of reproductive endocrine disruption studies (Markey et al., 2005; Peretz et al., 2014). Two adult females and one adult male were housed together in ventilated polysulfone cages until pregnancy was confirmed via the presence of a vaginal plug. Pregnant females were housed with one or two other females of the same dosing group until pregnancy day 17. On pregnancy day 18, each pregnant female was separated into her own cage in preparation for delivery. A 600–1,800 hours light cycle and a controlled temperature of 25–27°C was followed for the entirety of the experiment. Standard rodent chow (Harlan Teklad 2018, which has been reported to have minimal estrogenic activity (Vandenberg et al., 2008)) and tap water (in glass water bottles) were provided ad libitum. All procedures with mice were reviewed and approved by the Institutional Animal Care and Use Committee at the University of Massachusetts, Amherst.

4.2. Chemical administration

At the beginning of the dosing period, pregnant females were randomly assigned to exposure groups (n = 5 per group) using statistical methods to ensure similar body weights across treatment groups. Females were ear-tagged upon arrival to allow for their identification; identification tags were coded so that experimenters were blind to treatment throughout the experiment.

The day that a vaginal plug was observed was considered pregnancy day 1. Pregnant females were orally dosed with tocopherol-stripped corn oil (MP Biomedicals) alone (vehicle control, n = 5) or 200 μg BPS (Santa Cruz Biotechnology, 99% purity)/kg body weight/day in stripped corn oil (n = 5). Mice were trained to drink these substances (1 μL per gram body weight) from the end of a pipette. Pregnant and lactating mice were treated from pregnancy day 8 until lactational day 19. This period of perinatal exposure was selected to include the period of organogenesis as well as critical periods of differentiation and development for multiple organs including the brain; evaluation of other organs, body weight, and neurobehaviors have been published elsewhere (Catanese et al., 2015; Kim et al., 2015) or are forthcoming.

On postnatal day (PND) 19, two female offspring (F1 generation) were randomly selected from each litter, weaned from their mothers, and ear tagged. One female was fed tocopherol-stripped corn oil by pipette and the second was fed 1 μg/kg/day ethinyl estradiol (EE) in stripped corn oil. Both females were fed their assigned solutions from PND 19 to PND21 and were killed by cervical dislocation on PND22, 24 hours after the last dose of oil or EE. During this pre-pubertal period, female offspring were housed only with other females receiving the same postnatal treatment (oil or EE).

This dosing procedure produced four treatment groups: perinatal vehicle + prepubertal oil; perinatal vehicle + prepubertal EE; perinatal BPS + prepubertal oil; perinatal BPS + prepubertal EE (see Supplemental Figure 1). In all four groups, n = 5. This small sample size was selected for convenience rather than statistical power; this small sample size is a limitation of this pilot study.
4.3. Tissue and body weight collection

On PND22, female reproductive tissues including the ovary, oviduct and uterus were collected. All organs were examined for gross malformations including ovarian cysts. Uterine responsiveness to EE was confirmed visually via change in uterine size. One ovary and one uterine horn were snap-frozen in liquid nitrogen and stored at −80°C for qPCR analysis. One ovary, one oviduct, and one horn of the uterus were fixed in neutral buffered formalin for histological analyses. Body weight was also measured at time of sacrifice (data not shown).

4.4. Analysis of tissue morphology

Fixed tissues were processed through a series of alcohols and embedded in paraffin. Paraffin blocks were sectioned at 5 μm on a Fisher rotary microtome and mounted on positively-charged glass slides. Sections were processed through a series of xylene and ethanol, stained with hematoxylin and eosin, dehydrated, and coverslipped with permanent mounting medium. Images were collected on a Zeiss AxioImager Inverted Microscope with ZEN imaging software and a Zeiss high resolution color camera with a 20× EpiPlan Objective.

To analyze morphology of the oviduct and uterus, epithelial cell height was measured with ZEN imaging software. To score ovarian follicles, every tenth section throughout the entire ovary was analyzed, corresponding to one section per 50 μm of ovary. This method is not intended to count every follicle in the ovary; smaller follicles (primordial and primary) will be under-represented because of their small size. However, counting sections at this distance will prevent accidental counting of larger follicles more than once.

Follicles were scored using methods modified from those described previously (Hunt et al., 2012) and categorized depending on the stage of follicle development. Only follicles containing oocytes with visibly stained nuclei were counted. As expected, corpora lutea were not observed in any animal at PND22. To minimize statistical error due to variation between ovarian sizes, four sections (50 μm apart) from the medullary region were analyzed separately. For all tissues, sections were coded allowing for scoring by an observer blind to treatment group.

4.5. Immunohistochemistry

For immunohistochemical analysis of uterine, oviduct and ovarian tissues, standard protocols were used (Markey et al., 2005). Briefly, slides were processed through a series of xylene and ethanol washes followed by antigen retrieval using microwave-heated citric acid buffer. Endogenous peroxidase activity was quenched with hydrogen peroxide treatment and non-specific antibody sites were blocked using normal goat serum 1:20 in 1.5% milk. Ki67 antibodies were obtained from Vector Labs (rabbit polyclonal, Cat# VP-RM04) and used at concentrations of 1:1000. After overnight incubation with the primary antibody at 4°C, the tissue was incubated with the biotin labeled anti-rabbit secondary antibody (Abcam Cat# ab64256) for one hour at room temperature, and the reaction was developed with streptavidin peroxidase complex (Abcam Cat# ab64269). Finally, a color-change reaction was obtained using 3,3-diaminobenzidine (DAB, Abcam Cat# ab64238), tissue was counterstained with hematoxylin and cover-slipped with permanent mounting media.

To quantify Ki67 expression, one section of the uterus and the medullary region of the ovary were analyzed per female. Each sample was imaged with a Zeiss AxioImager Inverted Microscope with a 20× EpiPlan Objective with ZEN imaging software and a Zeiss high-resolution color camera. All analyses were conducted by an observer blind to treatment. In the ovary, follicles were analyzed by counting the number of positive and negative cells (both granulosa and theca cells were counted).
The total percentage of granulosa and theca cells that were positive for the specific antibody being tested was then calculated (# positive/total number counted × 100). Whenever possible, at least two follicles of the different stages were counted. In uterine and oviduct tissues, positive and negative cells were counted in the epithelial and stromal compartments, as appropriate.

4.6. TUNEL apoptosis assay

The Trevigen TACS 2 TdT-DAB *in situ* apoptosis detection kit was used for detection of apoptotic cells in ovarian and uterine tissue sections. Briefly, one section of the uterus and the medullary region of the ovary were analyzed per female. Slides were processed through a series of xylene and ethanol washes followed by 1× PBS. Samples incubated with Proteinase K solution for 15 minutes at room temperature, immersed in a hydrogen peroxide quenching solution and incubated with the reaction mix containing TdT enzyme and incubated for one hour at 37°C. The reaction was developed using a Strept-HRP and DAB solution. Samples were counterstained with hematoxylin, dehydrated, and cover-slipped with permanent mounting media.

Each sample was imaged with a Zeiss AxioImager Inverted Microscope with a 20× EpiPlan Objective with ZEN imaging software and a Zeiss high-resolution color camera. Samples were assessed using the same methods stated above for immunohistochemical analyses.

4.7. RNA isolation, cDNA synthesis and qPCR

Total RNA extracted from ovaries or uterine horns of individual mice using Trizol reagent (Ambion) was reverse transcribed using an iScript Reverse Transcription Supermix for RT-qPCR kit (Cat 170-884, BioRad). Primers were designed using free online software Primer3Plus and are listed in Table 2. The expression of four genes related to estrogen response was analyzed: Esr1 (ER α), Esr2 (ER β), IGF1 (insulin growth factor 1), and EGF-R (epidermal growth factor receptor). beta-2-microglobulin (B2 M) was used as a housekeeping gene; although this gene can respond to estradiol (Schröder, Pelch, & Nagel, 2009), we saw no changes in B2 M expression associated with BPS or EE treatment (data not shown). Triplicate 5-μL real-time PCR mixtures, each containing iTaq Universal SYBR Green Supermix (Cat 172-5124, BioRad), qPCR primers, and cDNA template were loaded onto a 384-well plate and run through 40 cycles on a CFX384 real time cycler (Bio-Rad Laboratories, Inc). Data were analyzed using the manufacturer’s CFX manager software, version 3.1. Relative quantification was determined using the ΔΔCq method to correct for differences in the reference gene (Livak & Schmittgen, 2001).

4.8. Statistical analysis

All statistical analysis was carried out with the use of SPSS statistical software version 22. Effect of chemical treatment was assessed for using independent sample t-tests to compare control and BPS groups, or oil and EE groups, as appropriate. Graphs indicate mean ± SEM. Results were considered significant at *p* < 0.05.

| Table 2. Primer sequences used for qPCR analysis |
|-----------------------------------------------|
| **Gene**                                      | **Forward primer sequence** | **Reverse primer sequence** |
| Housekeeping gene: beta-2-microglobulin/B2M | CCG GCC TGT ATG CTA TCC AG | TGT TCG GCT TCC CAT TCT CC |
| Epidermal growth factor receptor/EGF-R        | TCT TCA AGG ATG TGA AGT GTG | TGT ACG CTT TCG AAC AAT GT |
| Estrogen Receptor/ESR1                        | TGC AAT GAC TAT GCC TCT GG | CTC CGG TTC TTG TCA ATG GT |
| Estrogen Receptor/ESR2                        | ACT GCC AAT CAT CGC TTC TC | AGT AAC AGG GCT GCC ACA AC |
| Insulin like Growth Factor 1/IGF1             | GGA CCA GAG ACC CTT TGC GGG G | GGC TGC TTT TGT AGG CTT CAG TGG |
Abbreviations

BPA    bisphenol A
BPS    bisphenol S
DAB    3,3-diaminobenzidine
DES    diethylstilbestrol
EDCs   endocrine disrupting chemicals
EE     ethinyl estradiol
ER     estrogen receptor
NOAEL  no observed adverse effect level
PBS    phosphate buffered saline
PND    postnatal day

Supplementary material

The supplementary material for this paper is available online at https://doi.org/10.1080/2331205X.2017.1317690.

Declaration

The content of this manuscript is solely the responsibility of the authors and does not necessarily represent the official views of the National Institutes of Health or the University of Massachusetts.

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Competing interest

LNV has received travel reimbursement from Universities, Governments, NGOs and Industry, to speak about endocrine-disrupting chemicals. CEP, SS and AS have nothing to disclose.

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