Native Seed Supply and the Restoration Species Pool

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Abstract
Globally, annual expenditure on ecological restoration of degraded areas for habitat improvement and biodiversity conservation is approximately $18bn. Seed farming of native plant species is crucial to meet restoration goals, but may be stymied by the disconnection of academic research in seed science and the lack of effective policies that regulate native seed production/supply. To illustrate this problem, we identified 1,122 plant species important for European grasslands of conservation concern and found that only 32% have both fundamental seed germination data available and can be purchased as seed. The “restoration species pool,” or set of species available in practice, acts as a significant biodiversity selection filter for species use in restoration projects. For improvement, we propose: (1) substantial expansion of research and development on native seed quality, viability, and production; (2) open-source knowledge transfer between sectors; and (3) creation of supportive policy intended to stimulate demand for biodiverse seed.

Introduction
One-tenth of global wilderness has been destroyed in the last two decades (Pennisi 2016), and two-thirds of terrestrial environments are officially classed as degraded (Merritt & Dixon 2011). Ecological restoration (ER) accelerates the recovery of a degraded ecosystem with respect to health, integrity, and sustainability (SER 2004), and is recognized as a key complementary action for habitat conservation. Current global ER targets aim to restore 150 million ha or 15% of degraded ecosystems by 2020 (Menz et al. 2013). The estimated $18bn/year restoration cost is far exceeded by the potential global ecosystem service benefits of $85bn annually (Menz et al. 2013). Critical to success is the urgent need for access to high-quality seed through the farming of native species, as part of a range of flexible strategies to improve ER (Broadhurst et al. 2016).

Several large-scale ER initiatives are underway globally, such as the Australian Gondwana Link (Merritt & Dixon 2011), the Bureau of Land Management U.S. initiatives (Oldfield & Olwell 2015), the African Great Green Wall (Sacande & Berrahmouni 2016), and the European Union (EU) Natura 2000 (European Commission 1992). Seed-based plant conservation and use strategies (Merritt & Dixon 2011; Royal Botanic Gardens Kew 2015), seed-based research (Jiménez-Alfaro et al. 2016), and seed supply all play critical roles in successful ER. However, native seed sourcing, collection, production, and storage is more challenging than for agricultural species (Bischoff et al. 2016).
2008; Broadhurst et al. 2008) for which cultivars have been bred to be stable, uniform, and distinct (European Commission 1966).

ER depends on selecting appropriate species to cope with abiotic and biotic characteristics of degraded habitats. In ecological communities, scientists describe the species pool as the set of species that potentially occur at a site (Zobel 1992). The conditions limiting or facilitating species assembly will determine successional and recovering legacies of a system, including responses following ER (Temperton et al. 2004). Hand-collecting seed in large quantities from a broad range of species is unrealistic for most ER projects and wild populations risk depletion. Often the material used is restricted to that available from commercial or institutional seed suppliers. The “restoration species pool” (“RSP”), or pool of species available from these seed suppliers, thus imposes a critical biodiversity filter in ER projects. Where native supply lacks, easily available agronomic or horticultural seeds are used as a substitute, which is ecologically unacceptable. A RSP of native species, which has been systematically sourced between and within populations and species distribution ranges, is necessary for the support of genetic diversity in seed supplies and restored ecosystems (Hoban & Schlarbaum 2014).

Seed yields and germination of wild species can be naturally low and variable (Fenner 2000), and while cropping of native species can facilitate controlled production, some seed ecological traits (Fenner & Thompson 2005) can determine obstacles to harvesting. Not all wild species are candidates for commercial production as variation in seed morphological traits necessitates the use of appropriate harvesting and conditioning equipment, the costs of which can be very high if a large number of species are being produced. Proper seed management from collection to postconditioning storage is essential to maintain seed viability, which is variable between suppliers and can be very low (Marin et al. 2017). These challenges require collaborative efforts between seed suppliers and researchers to fully realize the potential of providing native farmed seeds for ER. This encompasses research on seed germination, dormancy (a process that regulates germination so that plants emerge under environmental conditions favourable for seedling establishment; Table S1), seed traits relevant for ER (Jiménez-Alfaro et al. 2016), and other bottlenecks that can be encountered such as adaptations for cultivation or genetic diversity maintenance (Chivers et al. 2016). However, research findings are rarely accessible to public stakeholders involved in ER.

Here, we assess the potential of the RSP to meet conservation needs in European grasslands, which are priority habitats as detailed in European policies on nature conservation. Human-induced habitat loss has impacted grassland biomes to the greatest rate and extent, largely due to agricultural conversion and the lack of conservation protections (Hoekstra et al. 2005). This neglect is in stark contrast to the biodiversity value of temperate grassland habitats, which across continental Europe are global biodiversity hot spots (Wilson et al. 2012). Using European grasslands of conservation concern as a case study, we analyze how many species have both detailed seed quality data and commercial seed lots available across taxa and across three species groups of relevance to European policies on ER. Addressing the availability of seed and related scientific information is important for the design of effective policy, research agendas, the foci of commercial seed suppliers, and reducing the risk of falling short in reinstating functional ecosystems in ER (Menz et al. 2013).

**Methods**

**Study systems and target species**

The European initiative Natura 2000 aims to establish a network of diverse, representative high-quality protected habitats of conservation concern, much of which will require intensive ER (European Commission 1992). Our study is focused on six major temperate grassland habitat types of conservation concern in Europe: lowland meadows (Natura 2000 number: 6510); high altitude hay meadows (6520); dry grasslands (6210); species rich Nardus grasslands (6230); calcareous alpine grasslands (6170); and acidic alpine grasslands (6150).

We created a database of 1,122 target species with potential interest for ER within these habitats, regulated by EU legislation that affects strategies of seed quality and use (Table S2). This includes 116 protected species subjected to legal protection, in most cases endangered or narrow endemic species; 929 indicator species, which are indirectly protected when occurring in protected habitats but unregulated in seed production; and 77 fodder species controlled for quality as domestic stock feed (European Commission 1966; 2014), as well as for preservation of genetic diversity (European Commission 2010; Table 1).

To assess the availability of seed quality data, we collected trait information on germination temperature and dormancy type of the target species available from the Seed Information Database (Royal Botanic Gardens Kew 2008), and the most recent review of seed germination studies (Baskin & Baskin 2014). As these are the main traits related to the germinability of a seed lot, we assume that having this information implies a minimum contribution of the scientific community for a given
Table 1 Relevant legislation details related to each target species group

| Species group | Description | Legislation | Impact |
|---------------|-------------|-------------|--------|
| Protected species (N = 116) | Includes species of conservation concern, in most cases endangered or narrow endemics, listed by name in relevant policy, and occurring in focus habitats. | Specific species for which member states must protect and conserve when found to occur under Annex II & IV of the EU policy on Conservation of Natural Habitats Wild Fauna and Flora (European Commission 1992). | Species seed cannot be collected without a rigorous permit process. |
| Indicator species (N = 929) | Species that are diagnostic or dominant for any of the selected habitats at the continental scale according to Schaminée et al. (2016) and vegetation ecology literature (Georg Grabherr & Mucina 1993). | These species are indirectly conserved in Annex II as reflected in the designation of special protected areas for the habitats in that they occur under the EU policy on Conservation of Natural Habitats Wild Fauna and Flora (European Commission 1992). | Species are of interest for use in restoration and have no direct EU policy restrictions on their collection, reproduction, or use but may have local regulations. |
| Fodder species (N = 77) | Grass and legume species used for animal forage, also considered valuable for preservation of the natural environment and conservation of genetic resources in grasslands listed by name under relevant policies. | Specific species and genera important for domestic stock and grazing (European Commission 1966, 2010, 2014). | Controlled for quality including high purity standards and minimum germination thresholds in EU Commission Directive 1966. Expanded in Directive (2010) to include harvest method, seed weight, quantity, region of origin, source area (collection site and multiplication), habitat type, and year of collection. Native seed production cannot exceed 5% of the total commercial cultivar production market in their country. |

N = number of species in each group.

species. A systematic online search was conducted from November 2014 to May 2016, and the lists of species available commercially as seed were downloaded, or requested to seed suppliers. As there are multiple seed sources in some countries, the supplier providing the highest number of target species was selected since the inclusion of smaller companies did not influence the total number of available species. This resulted in seed availability lists from 17 seed suppliers across 17 countries (Table S3). Species names were verified against The Plant List (Missouri Botanical Gardens, Royal Botanic Gardens Kew 2013). Possible limitations of these data are that species reported as available may be an overestimate as lists may be outdated, inaccurate, or in some cases represent cultivars rather than native species, particularly in the fodder group. Nonetheless, the list is an accurate representation of the current state of native seed acquisition in Europe. We use the term supplier instead of producer because in the majority of cases, seed is reproduced in a native seed farm or orchard, but in some cases seed may be hand-collected.

Analyses

Data were collected as binomial variables. To assess Germination Data Availability (GDA), each species was assigned as data being available (1) or not (0). Similarly, species were either Commercially Available (CA) (1) or not (0).

The proportions (%) of species with CA and with GDA were calculated for each plant family represented in the target species list to elucidate taxonomic representation as a surrogate of phylogenetic variation. A Generalized Linear Model (GLM) was fitted to assess the variation of CA as a function of GDA and species groups. The GLM was computed with binomial error distribution and logit link function in order to assess the influence of policy groups and GDA (explanatory variables) on CA (response variable; CA ~ GDA + policy group). All analyses were performed in R Statistical Computing Language and Platform version 3.2.2 (R Core Development Team 2016), and Figures created in the package ggplot2 (Wickham 2009) and yarrr (Phillips 2016). The package Effects (Fox 2003) was used to create probability estimates of CA based on each variable.
The restoration species pool

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Figure 1  (A) Proportion (%) of species that are commercially available (CA) and with germination data availability (GDA) (B) proportion (%) of species that are commercially available (CA) with germination data availability (GDA), and with the combination of CA + GDA. N: number of species represented within each group.

and package PMCMR for post hoc pairwise Kruskal-Wallis tests (Pohlert 2014).

Results

The 1,122 target species with potential interest for ER within European grassland habitats are spread across 59 plant families, with highest representation in Compositae (146 species) and with the top 5 and 10 families comprising 43% and 62% of the species list, respectively. Information on GDA and CA alone extended to 49% (i.e., 556) and 39% (i.e., 439) of target species, respectively (Figure 1A). Information for both seed GDA and CA details are available for only 32% (i.e., 358) of species on the target list (Figure 1B). Supplied seed is not available across all suppliers (Figure S1), although indicator and fodder species with GDA are available across a higher proportion of suppliers than those without GDA and with protected status (Kruskal-Wallis $s^2 = 338.81$, $P \leq 0.001$; Tables S2 and S4).

The majority of taxonomic families completely lacking GDA are also completely lacking CA, although the sample size is small in these cases (Figure 2, Table S5). The vast majority of families with large sample sizes have ~50% GDA and CA. Within this case study, there are seven families, spanning nine genera and 15 species, for which germination data are unknown. Twelve families (20% of total) lie within the lower quartile of CA, covering 158 species (14% of total).

Strong predictive patterns based on the GLM are exhibited for the estimate of CA of target species across all variables (Figure 3, Table 2). The model predicts that protected species have a 0.04 probability of being CA, indicator species 0.37 ($P < 0.001$), and fodder species 0.54 ($P < 0.001$; Figure 3A, Table 2). Species with no GDA have 0.13 probability of being CA, and species with GDA have a 0.58 probability of being CA overall ($P < 0.001$; Figure 3B). The combination of predictors (Figure 3B) provides a further level of outcomes. Protected species for which there is no GDA have 0.01 probability of CA; this probability increases to 0.11 when there is GDA. Comparable values for indicator species without and with GDA are 0.17 and 0.64, respectively; and 0.29 and 0.78 probability, respectively, for fodder species.
Figure 2. Bars show the proportion (%) of species per taxonomic family that have seed which has commercial availability. The degree and proportion of germination data availability is represented by the color scale according to the Seed Information Database (Royal Botanic Gardens, Kew 2008) and Baskin & Baskin (2014). The numbers in brackets next to each family name represents how many species are included in the data set from that given family.

**Discussion**

**The Restoration Species Pool in European grasslands**

To our knowledge, no studies have investigated the availability of commercial seed and related germination data for native seed in a large-scale case study. In Europe, the relatively high availability of native seeds for *fodder* species demonstrates that commercial availability of native seed is subject to economic demand and a long-standing regulatory framework. This framework follows an agricultural model meant for animal feed rather than Ecological Restoration (ER) (European Commission 1966), yet is recognized for ER use (European Commission 1966).
The restoration species pool

Probability estimate of commercial availability

Protected Indicator Fodder

A

B

Figure 3 Predicted effect plots showing the commercial availability of species grouped per species category. Probability was estimated using GLM (binomial error, logit link) fitted to the commercial production data of each species (commercial availability ~ germination data availability + species group). The same model was used to fit each group, and results were grouped based on: (A) species groups (protected, indicator, fodder) (B) species group + germination data availability.

Bars represent the probability that a given group of species is commercially available. Brackets represent the upper and lower limits of that estimate.

N = number of species represented by each prediction.

Table 2 Generalized linear model (binomial error, logit link) analysis testing the effect of germination data availability and species group on commercial availability

| Coefficient                     | Effect estimate | Standard error | Z      | P       |
|---------------------------------|-----------------|----------------|--------|---------|
| Intercept (protected)           | −4.2541         | 0.6016         | −7.071 | < 0.001 |
| Germination data availability   | 2.1759           | 0.1524         | 14.281 | < 0.001 |
| Indicator species               | 3.3685           | 0.6570         | 5.127  | < 0.001 |
| Fodder species                  | 2.6502           | 0.6021         | 4.401  | < 0.001 |

Results were estimated using GLM fitted to the commercial availability data for each species.
and storage strategies. In Australia, United Kingdom, and the United States, there are examples of government, community, or nonprofit groups working cooperatively with seed suppliers to enable the inclusion of species that have challenging seed traits in the commercial RSP supply chain. The U.S. Native Plant Program (Oldfield & Olwell 2015) contracts production of seed across all available suppliers, to partition demand and market share, then stored in government infrastructure for purchase. As a unique example in Europe, Germany has mandated that only native species may be used for all revegetation by 2020 (BNatSchG, Federal Ministry for the Environment, Nature Conservation, and Nuclear Safety 2010). Compared to German native seed demand in 2015, production of local native seeds must grow 10-fold to meet 2020 targets (Pers. Comm., Ann Kareen Mainz, Association of German Wild Seed Producers), an increase which will require expansion of their RSP. Demand creation, contracting, storage, and provision solutions must be developed in tandem to effectively expand RSP capacity.

Policy recommendations

Current legislation relating to protected (European Commission 1992) and fodder (European Commission 2010) species recognize the need to produce seed specifically for ER, but do not match the native seed market appropriately. Policy relating to the use of seed mixtures mandates that commercially produced seed must come from the same source area in which it is being used, and germination minimums are required (European Commission 2010), which are easily achievable in cultivars, but unrealistic for native species. These quality standards are too restrictive (Tishew et al. 2011), to which there is low adherence and enforcement, as they are contradictory to a much-needed industry with a small market niche. Supportive regulation is needed and future EU policy should require that all public revegetation projects use only native material. Creating demand through policy while aligning the contracting of supply offers immense potential to enable growth of the RSP. We strongly support initiation of policies to contract annual native seed production of baseline indicator and fodder species across available producers to store for large-scale projects. Policy should require vegetation biodiversity targets to be met in ER and revegetation. Sourcing and contracting of site-specific seed material beyond yearly indicator and fodder stores (including but not limited to protected species) should be required at project inception to allow time for realistic production. New policies should be designed to embrace consultation with the native seed industry and restoration professionals.

Conservation seed banks for native species can support these strategies in a small capacity and can provide access to relevant small-scale seed processing and quality assessment equipment (Nevill et al. 2016). The largest ex situ plant conservation programme globally, the Millennium Seed Bank Partnership (MSBP), managed by the Royal Botanic Gardens, Kew, UK, has successfully banked seeds of 13% of the world’s wild species, aiming to bank 25% by the year 2020 (Royal Botanic Gardens Kew 2015). Seed from the MSBP has been used for small-scale re-establishment, generally targeted for threatened species. An exemplar is FAO-RBG Kew “Africa’s Great Green Wall” program within which collaborating country seed banks supply ~25,000 kg of seed per annum of about 200 species of trees, shrubs, and grasses (Sacandé & Berrahmouni 2016). Nevertheless, a new form of Restoration Seed Banks (Merritt & Dixon 2011) is needed if a sustainable seed supply chain of the right scale is to be supported for the RSP. To improve ER outcomes, wide expansion of current capacity and collaboration across sectors is needed to provide the requisite tons of native seed needed (Merritt & Dixon 2011). In addition, research in seed biology and vegetation science applied to seed sourcing, applications, and bottlenecks related to collection and use are required.

Current research in seed biology and regeneration processes remains specialized, in need of urgent expansion (Larson & Funk 2016). In addition, long-term interdisciplinary and collaborative open-source knowledge sharing platforms are needed to facilitate the exchange of research (Royal Botanic Gardens Kew 2015). We suggest future germination research focus on the development of efficient dormancy breaking treatments, the thermal control of germination (thresholds and rates), and improvements in native seed production practices for European grassland species not currently covered by the RSP. Integration of research and industry knowledge sharing where any research project connected to native seed germination delivers findings to the private sector could hold wide benefits. Research projects for protected or underrepresented taxa could ideally include commercial or cooperative seed production contracts for direct use in conservation and reintroduction as industry output components. Supplying protected species must be strictly designed, implemented, and controlled with the direct use of vanguard science through extremely collaborative approaches (Shirey et al. 2013).

Conclusions

Our analysis presents the first study investigating seed germination data availability and the commercial “RSP.” We present a continental case study, reflecting a global issue of global importance to habitat conservation. In
summary, we encourage further exploration and reconsideration of public policy, compilation of open-access knowledge sharing across sectors and multinational efforts to provide infrastructure and support, so as to expand and realise the full potential of the emerging native seed industry. Improving the breadth of seed biology research and knowledge sharing between sectors has potential to support the expansion of the commercial native seed market and the RSP. Improved commercial availability could reduce species bias and risk in ER.

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Supporting Information

Additional Supporting Information may be found in the online version of this article at the publisher’s web site:

Figure S1 Observed percentage (%) of suppliers (total 17) with commercial availability of seed with and without germination data availability.

RDI plots (raw data, descriptive and inference statistics) show jittered points of raw data, center bars indicate the mean of the data, beans outline the smoothed density of the data, whiskers mark the 10% and 90% quantiles of the data, and inference bands show the Bayesian 95% high-density interval inferential statistics for each group. Letters show statistical differences between groups (Table S4).

Table S1 Simplified seed dormancy types (adapted from Baskin & Baskin 2014)

Table S2 Full species list, associated category, and associated data

CA = commercial availability (yes [1], no [0]), GDA = germination data availability (yes [1], no [0]).

Table S3 Seventeen seed suppliers across 17 countries used for data collection

Statistics representing differences between variables in the percentage of suppliers with seed of each species commercially available compared across species groups (Figure S1)

Kruskal-Wallis $\chi^2$ test and post hoc pairwise Tukey and Kramer (Nemenyi) $\chi^2$ test, $P$-value statistics, indicating significance between group variables. Germination data available $= + \text{GDA},$ germination data not available $= - \text{GDA}.$

Table S5 The complete data set summarized by taxonomic family in descending order of percentage of commercial availability (CA)

# = number, % = percentage, Sp. = species, CA = commercially availability, GDA = germination data availability.

References

Baskin, C.C. & Baskin, J.M. (2014). Seeds: ecology, biogeography, and evolution of dormancy and germination. 2nd edn. Elsevier, San Diego, London, Waltham.

Bischoff, A., Steinger, T. & Muller, S.Á.H. (2008). The importance of plant provenance and genotypic diversity of seed material used for ecological restoration. Restor. Ecol., 18, 338-348.

Broadhurst, L.M., Lowe, A., Coates, D.J., et al. (2008). Seed supply for broadscale restoration: maximizing evolutionary potential. Evol. Appl., 1, 587-597.

Broadhurst, L.M., Jones, T.A., Smith, F.S., North, T. & Guja, L. (2016). Maximizing seed resources for restoration in an uncertain future. BioScience, 66, 73-79.

Chivers, I.H., Jones, T.A., Broadhurst, L.M., Mott, I.W. & Larson, S.R. (2016). The merits of artificial selection for the development of restoration-ready plant materials of native perennial grasses. Restor. Ecol., 24, 174-183.

European Commission. (1966). Council Directive 66/401/EEC of 14 June 1966 on the marketing of fodder plant seed.

European Commission. (1992). Council Directive 92/43/EEC of 21 of May 1992 on the conservation of natural habitats and wild fauna and flora.

European Commission. (2010). Commission Directive 2010/60/EU of 30 August 2010 providing for certain derogations for marketing of fodder plant seed mixtures intended for use in the preservation of the natural environment. Text with EEA relevance. Official Journal of The European Union.

European Commission. (2014). Commission Implementing Decision 2014/362/EU of 13 June 2014 amending Decision 2009/109/EC on the organisation of a temporary experiment providing for certain derogations for the marketing of seed mixtures intended for use as fodder plants pursuant to Council Directive 66/401/EEC.

Federal Ministry for the Environment, Nature Conservation, and Nuclear Safety. (2010). Act on Nature Conservation and Landscape Management (Federal Nature Conservation Act – BNatSchG) of 29 July 2009. Federal Nature Conservation Act. 1.
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Fenner, M. (2000). Seeds: the ecology of regeneration in plant communities. 2nd edn. CABI Publishing, Wallingford, New York.

Fenner, M. & Thompson, K. (2005). The ecology of seeds. Cambridge University Press, Cambridge.

Fox, J. (2003). Effect displays in R for generalised linear models. J. Stat. Soft., 8, 1-27.

Georg Grabherr, von H. & Mucina, L. (1993). Die Pflanzengesellschaften Osterreichs. Gustav Fischer Verlag Jena, New York.

Hoban, S. & Schlarbaum, S. (2014). Optimal sampling of seeds from plant populations for ex-situ conservation of genetic biodiversity, considering realistic population structure. Biol. Conserv., 177, 90-99.

Hoekstra, J.M., Boucher, T.M., Ricketts, T.H. & Roberts, C. (2005). Confronting a biome crisis: global disparities of habitat loss and protection. Ecol. Lett., 8, 23-29.

Jiménez-Alfaro, B., Silveira, F.A.O., Fidelis, A., Poschlod, P. & Commander, L.E. (2016). Seed germination traits can contribute better to plant community ecology. J. Veg. Sci., 27, 637-645.

Larson, J.E. & Funk, J.L. (2016). Regeneration: an overlooked aspect of trait-based plant community assembly models. J. Ecol., 104, 1284-1298.

Marin, M., Toorop, P., Powell, A.A. & Laverack, G. (2017). Tetrazolium staining predicts germination of commercial seed lots of European native species differing in seed quality. Seed Sci. Technol., 45, 151-166.

Menz, M.H.M., Dixon, K.W. & Hobbs, R.J. (2013). Hurdles and opportunities for landscape-scale restoration. Science, 339, 526-527.

Merritt, D.J. & Dixon, K.W. (2011). Restoration seed banks - a matter of scale. Science, 332, 424-425.

Missouri Botanical Gardens, Royal Botanic Gardens Kew. (2013). The plant list. http://www.theplantlist.org/ (visited May 4, 2016).

Nevill, P.G., Tomlinson, S., Elliott, C.P., Espeland, E.K., Dixon, K.W. & Merritt, D.J. (2016). Seed production areas for the global restoration challenge. Ecol. Evol., 6, 7490-7497.

Oldfield, S. & Olwell, P. (2015). The right seed in the right place at the right time. BioScience, 65, 955-956.

Pennisi, E. (2016). We’ve destroyed one-tenth of Earth’s wilderness in just two decades. Science. http://www.science mag.org/news/2016/09/we-ve-destroyed-one-tenth-earth -s-wilderness-just-2-decades?utm_source=sciencemagazine &rutm_medium=facebook-text&utm_campaign=wilder ness-7205 (visited Oct. 15, 2016).

Phillips, N. (2016). yarr: a companion to the e-book “YaRrr!: the pirates guide to R.” http://nathanielpillips.com/ thepiratesguidetor/ (visited Aug. 5, 2016).

Pohler, T. (2014). The pairwise multiple comparison of mean ranks package (PMCMR). http://CRAN.R-project.org/ package=PMCMR (visited Sept. 2, 2016).

R Core Development Team. (2016). R: a language and environment for statistical computing. https://cran.r-project.org/ (visited September 2, 2016)

Royal Botanic Gardens Kew. (2008). KEW seed information database (SID). http://data.kew.org/sid/ (visited May 4, 2016). http://www.kew.org/science/who-we-are-and-what-we-do/kews-science-strategy (visited Sept. 2 2017).

Royal Botanic Gardens Kew. (2015). Kew science strategy 2015–2020.

Sacande, M. & Berrahmouni, N. (2016). Community participation and ecological criteria for selecting species and restoring natural capital with native species in the Sahel. Restor. Ecol., 24, 479-488.

Schaminée, J.H.J., Chytrý, M., Hennекens, S.M., Janssen, J.A.M., Jiménez-Alfaro, B., Knollóvá, I., Marceno, C., Mucina, L., Rodwell, J.S., Tichý L. & data-providers. (2016). Review of grassland habitats and development of distribution maps of heathland, scrub and tundra habitats of EUNIS habitats classification. Report EEA/NSV/15/005. European Environment Agency, Copenhagen.

Shirey, P.D., Kunicky, B.N., Chaloner, D.T., Bruescke, M.A. & Lambertii, G.A. (2013). Commercial trade of federally listed threatened and endangered plants in the United States. Conserv. Lett., 6, 301-316.

Society for Ecological Restoration, Science and Policy Working Group (SER) (2004). SER international primer on ecological restoration. Society for Ecological Restoration International. http://www.ser.org/page/SERDocuments (visited Oct. 4, 2016).

Suding, K.N., Gross, K.L. & Houseman, G.R. (2004). Alternative states and positive feedbacks in restoration ecology. Trends Ecol. Evol., 19, 46-53.

Temperton, V.M., Hobbs, R.J., Nuttle, T. & Halle, S. (2004). Assembly rules and restoration ecology. Island Press, Washington, Covelo, London.

Tishew, S., Youtie, B., Kirmer, A. & Shaw, N. (2011). Farming for restoration: building bridges for native seeds. Ecol. Restor., 29, 219-222.

Wickham, H. (2009). ggplot2: elegant graphics for data analysis. Springer-Verlag, New York.

Wilson, J.B., Peet, R.K., Dengler, J. & Partel, M. (2012). Plant species richness: the world records. J. Veg. Sci., 23, 796-802.

Zobel, M. (1992). Plant-species coexistence - the role of historical, evolutionary and ecological factors. Oikos, 65, 314-320.