Diesel, children and respiratory disease

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ABSTRACT

Air pollution generated in urban areas is a global public health burden since half of the world’s population live in either cities, megacities or perurban areas. Its direct effects include initiating and exacerbating disease, with indirect effects on health mediated via climate change putting the basic needs of water, air and food at risk.

INTRODUCTION

There is strong epidemiological evidence that air pollution is associated with a wide range of adverse health effects on the respiratory, cardiovascular and neurological systems. Indeed, in the UK, the combination of new-onset (incident) diseases associated with long-term exposure, and exacerbation of diseases once disease is established results in approximately 40,000 excess deaths a year that are attributable to air pollution, increasing health service and social costs by over £20 billion a year. Although deaths associated with air pollution are mainly in adults, there is also increasing concern that air pollution, especially from diesel vehicles, has major adverse effects in children and that this has long-term consequences. In this review, we report the evidence that underpins the need for exposure reduction policy to focus on diesel vehicles and the potential beneficial effects of such a policy on children’s health. Although this review focuses on the heavily dieselised UK environment, it is also relevant to countries where diesel vehicles remain a major source of emissions.

COMPONENTS OF AIR POLLUTION

The major outdoor pollutants in urban areas are inhalable particulate matter (PM, measured as either PM less than 10 µm in aerodynamic diameter (PM$_{10}$) or the even smaller PM$_{2.5}$), nitrogen oxides (NO$_X$), such as nitrogen dioxide, NO$_2$, ozone (O$_3$), sulfur dioxide (SO$_2$), carbon monoxide (CO) and hydrocarbons (HC). Sources of these include gasoline-powered and diesel-powered engines from vehicles, trains and, in port towns, ships (proximately PM$_{10}$, vehicle tyre and brake wear (PM), power stations and factories from coal combustion and biomass burning (PM, NO$_X$ and SO$_2$), and wood burning heating that is increasingly popular, contributing up to 9% of PM in London during winter. For diesel engines, an important component of emissions is black carbon, that is, the fraction of PM that most strongly absorbs light—a component that is often called ‘diesel soot’. Another pollutant, ozone, is formed by the reaction of NO$_X$ with carbon compounds called volatile organic compounds (VOCs) in the presence of sunlight. Two of the most important VOCs emitted by vehicles are benzene and 1,3-butadiene. For emissions from diesel, there is a strong correlation between locally emitted PM$_{10}$ and NO$_X$. and it is reasonable to assume that, where diesel vehicles predominate, either metric is a good marker of exposure to the locally generated pollutant mix in urban areas.

WHY FOCUS ON DIESEL?

Many parts of the UK breach the EU legal limits and WHO guidelines (table 1) for pollutants on a regular basis. While London often exhibits the biggest breach of pollution limits, other parts of the UK are also affected. Indeed, a recent report from the Department
hydrocarbons (HC), PM and NOx, all of which are associated with negative health effects. The reason why diesel engines should be a major target for exposure-reduction strategies is that they emit more PM and NOx than their petrol or hybrid counterparts, contributing to about 40% of all NOx emissions in inner cities. Furthermore, diesel, and not petrol, soot is categorised by WHO as carcinogenic—a categorisation that implies that diesel PM is, mass for mass, more toxic than petrol PM. Vehicle emissions are regulated by the European Union (EU) Euro standards, currently at Euro 6 (table 2).

Compliance with Euro standards is assessed under laboratory conditions only and are less strict for diesel engines. But even given this leeway, recent measurements under real-life driving conditions have shown that diesel cars produce significantly more toxic emissions than the Euro standard, whereas petrol engines map closely to the laboratory Euro standard (figure 1); this phenomenon is observed globally, and Anenberg et al. reported approximately a third of heavy-duty and over half of light-duty diesel vehicle emissions breaching the certification limits, across 11 major vehicle markets. Thus, over 2000 education or childcare providers in England and Wales are located close to busy roads with concentrations of NOx that are regularly higher than legal limits (40 µg/m³ annual mean or 200 µg/m³ 1 hour mean). In addition, children attending these schools are exposed to high concentrations of freshly generated diesel pollutants during the commute to and from school and during outdoor activities (figure 2).

### HEALTH EFFECTS OF DIESEL EMISSIONS ON CHILDREN

Few epidemiological studies address the effects of diesel emissions alone. However, it is reasonable to extrapolate from studies that have assessed exposure to either PM or NOX since (1) diesel PM is not less toxic than other types of PM, and (2) the adverse effects of gases such as NOX are independent of source. One way of estimating the health burden from diesel emissions alone is to use source apportionment data. For example, in London where most taxis, buses, heavy goods vehicles and vans are powered by diesel (table 3), 48% of NOX and 54% of PM10 is from road transport (figures 3 and 4). These vehicles, along with diesel cars, are responsible for 34% of total PM10 and 38% of total NOX emissions (figures 3 and 4). When considering effects measured in later childhood, it is difficult to separate the effect of maternal exposure

### Table 1 EU limits, WHO guidelines and main sources of ambient (outdoor) air pollutants. Adapted from European Commission Air Quality Standards (updated September 2017), WHO Ambient (outdoor) air quality and health fact sheet (updated Sept 2016), and Lethal and Illegal, Solving London’s Air Pollution Crisis by Institute for Public Policy Research, November 2016

| Pollutants                  | EU legal limits (averaging period) | WHO guidelines (averaging period) | Main sources                                      |
|----------------------------|-----------------------------------|-----------------------------------|--------------------------------------------------|
| Nitrogen dioxide (NOx)     | 200 µg/m³ (1 hour) 40 µg/m³ (1 year) | 200 µg/m³ (1 hour) 40 µg/m³ (1 year) | Transport, combustion                             |
| Ozone (O3)                 | 120 µg/m³ (8 hours)               | 100 µg/m³ (8 hours)               | Reaction of hydrocarbons, nitrogen oxides and volatile organic compounds in sunlight |
| Particulate matter (PM10)  | 50 µg/m³ (24 hours) 20 µg/m³ (1 year) | 50 µg/m³ (24 hours) 20 µg/m³ (1 year) | Transport (exhaust, tyre, brake wear), combustion, industrial processes and construction |
| Particulate matter (PM2.5) | 25 µg/m³ (1 year)                 | 10 µg/m³ (24 hours) 25 µg/m³ (1 year) |                                                  |
| Sulfur dioxide (SO2)       | 350 µg/m³ (1 hour) 20 µg/m³ (24 hours) | 500 µg/m³ (10 min) 20 µg/m³ (24 hours) | Coal combustion and road transport                |

### Table 2 EU Euro emissions standards. Adapted from Lethal and Illegal, Solving London’s Air Pollution Crisis by Institute for Public Policy Research, September 2016, and SMMT Euro Standards for Cars (accessed March 2018)

| Euro emissions standards | NOx (g/km) | PM10 (g/km) | NOx (g/km) | PM10 (g/km) |
|-------------------------|------------|-------------|------------|-------------|
| Petrol cars              |            |             |            |             |
| Euro 4 (2005)            | 0.08       | –           | 0.25       | 0.025       |
| Euro 5 (2009)            | 0.06       | 0.005       | 0.18       | 0.005       |
| Euro 6 (2014)            | 0.06       | 0.005       | 0.08       | 0.0045      |
| Diesel cars              |            |             |            |             |

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to air pollution from postnatal effects—since there is a strong correlation between exposure to traffic-derived air pollutants (TRAPs) of pregnant women and their children. But independent associations between antenatal exposure to NO$_2$ and reduced FEV$_1$ later in childhood are reported. For example, Morales et al$^{24}$ reported that an IQR increase in NO$_2$ exposure during the second trimester was associated with an estimated change in childhood FEV$_1$ by −28 mL, while the relative risk of having FEV$_1$ <80% predicted was 1.30. By contrast, effects on the fetus or on the newborn infant must be due to maternal exposure. These epidemiological studies report
that maternal exposure to TRAP has adverse effects on the fetus leading to increased infant mortality, reduced fetal growth, low birth weight at term and premature birth. Indeed, increased risk for the low birth weight for term metric is found at maternal PM$_{2.5}$ exposure lower than the EU recommended annual limit of 25 µg/m$^3$. It is likely that these antenatal effects synergise with postnatal pollution exposures to increase susceptibility to common respiratory conditions such as wheeze, bronchiolitis and asthma.

**Childhood exposure**

Air pollutants, particularly NO$_X$ (reflecting exposure to both NO$_X$ and PM), are associated with reduced lung function in children—for both FVC and FEV$_1$. Urman et al$^5$ showed that an increase of 17.9 ppb of NO$_X$ exposure was associated with a 1.56% deficit in FVC and 1.1% deficit in FEV$_1$, and similar findings were seen in children with or without asthma. Residing in areas with high concentrations of PM and NO$_2$ can also lead to suppression of lung function growth in school children$^{4,31}$.

This reduction can potentially be halted and reversed with better air quality. For example, Gauderman et al$^{32}$ showed that reducing the levels of NO$_2$, PM$_{10}$ and PM$_{2.5}$ were associated with improvements in FEV$_1$ and FVC growth in adolescents over 4 years—mean 4-year growth in FEV, increased by 91.4 mL per 14.1 ppb of NO$_2$ reduction, and 65.5 mL per 8.7 µg/m$^3$ of PM$_{10}$ reduction, and 65.5 mL per 12.6 µg/m$^3$ of PM$_{2.5}$ reduction, with comparable findings in FVC. Children with existing chronic illnesses, particularly respiratory conditions, are most vulnerable.

Air pollution can predispose individuals to new-onset asthma; preschool children are more prone to new onset of wheeze. A meta-analysis concluded that exposure to NO$_2$ is linked to new-onset asthma, while exposure to PM is linked to new-onset wheeze.$^{33}$ An effect of diesel PM per se on reactivity to inhaled allergens is supported by the association between long-term traffic pollution exposure and allergies.$^{34-36}$ Asthma exacerbations are also closely associated with short-term variations in PM$_{2.5}$. Although increasing inhaled corticosteroids prior to high pollution days may seem logical,$^{38}$ it is unclear whether this strategy is effective.

There is emerging evidence that air pollution impacts on children’s neurological system and development. For example, associations between exposure to air pollutants and reduced IQ and neurocognitive ability such as working memory, autism and reduced brain-derived neurotrophic factor are widely reported.$^{39-41}$ In particular, Basagaña et al$^{39}$ reported that traffic-related PM$_{2.5}$ was more strongly associated with reduction in cognitive function compared with fine particulates from other sources.
sources such as mineral, heavy oil combustion or road dust. In addition, exposure to high levels of traffic-induced pollutants may delay maturation of the brain. An additional emerging link is between air pollution and the endocrine system. For example, Thiering et al reported an association between insulin resistance and either NO\textsubscript{2} or PM exposure in healthy children.

**Implications for adult life**

It is increasingly recognised that impaired fetal well-being is a substrate for adult-onset cardiovascular disease such as atherosclerosis. Prolonged exposure to air pollutants may increase mean pulmonary arterial pressure and diastolic blood pressure, predisposing to cardiovascular events and premature death in adulthood. The effect on cognition lingers onto adulthood, where associations with dementia and Parkinson’s disease have been found.

Although the epidemiological evidence for the health effect of fossil-fuel-derived pollution is very strong, there are important confounders that must be considered. For example, in England, increased exposure to mean annual NO\textsubscript{2} concentrations is higher in areas of increased social deprivation and reduced access to healthcare. Furthermore, children from more deprived areas are also more likely to be exposed to other sources of pollution such as second-hand cigarette smoking.

**Mechanisms**

Many of the mechanisms underlying the robust epidemiological associations between air pollution and health across the life course remain to be defined. Effects on organs distant from the lung are likely to be facilitated by mediators released in the airway subsequently leaching out into the systemic circulation. A key cell for release of mediators is alveolar macrophage (AM) since phagocytosis of PM by AM stimulates release of cytokines such as interleukin-6, interleukin-8 and tumour necrosis factor. PM that reaches the most distal airways is phagocytosed in a dose-dependent manner by airway macrophages (figure 5). Indeed, Kulkarni et al reported that in healthy children, the amount of carbon in AM (as a marker of long-term personal exposure) is inversely associated with lung function. Phagocytosis of inhaled diesel PM by AM is also essential for normal removal of PM from the lungs, which minimises exposure of other airway cells. Conditions that impair AM phagocytosis will increase the proportion of PM impacting on and penetrating airway epithelial cells, further worsening the release of inflammatory mediators. Indeed, a recent study found significantly lower amounts of diesel soot in AM, compatible with abnormal clearance of inhaled PM, in children with moderate-to-severe asthma compared with healthy controls—despite similar levels of personal exposure to black carbon.
Increased exposure of airway epithelial cells to PM increases the potential for PM to translocate into the systemic circulation and directly cause adverse effects in distant organs, including the fetus where transplacental transfer of nanomaterials up to 240 nm is possible.

WHAT CAN WE DO ABOUT DIESEL POLLUTION?

National level

In London, air pollution is mostly caused by road traffic, of which diesel vehicles are a major contributor, as discussed above. With an estimated 9400 premature deaths attributable to air pollution, it has the second biggest impact on public health. These highly polluting vehicles should therefore be phased out to comply with legal limits of pollutants—and cleaner alternatives encouraged. Tougher national regulations on traffic emissions such as the expansion of Ultra Low Emission Zones and scrappage schemes for older generations of diesel vehicles should be considered. Indeed, the 2016 report from the Institute for Public Policy Research estimated that phasing out diesel-powered vehicles in London would lead to large reductions in NO$_x$ and NO$_2$ levels, ultimately lowering NO$_2$ levels to comply with EU standards. This report estimated that with a 45% reduction in NO$_x$ and 56% reduction in NO$_2$, 1.4 million life-years would be gained along with a financial benefit of up to £800 million.

Planting trees can reduce air pollution by acting as a physical barrier to intercept PM and absorbing gaseous pollutants such as O$_3$, although the effect on pollution concentrations at schools is, to date, unclear. However, the amount of pollutants removed by these organic barriers will be proportional to the extent of plantation.

Therefore, vast tall hedges around nurseries and schools should be encouraged, but this does not provide protection against pollution exposure during travel to and from schools.

Individual level

Various measures such as walking along less busy roads, cycling, use of public transport and carpooling may reduce exposure to air pollution, but the evidence base for whether this is achievable over the long term, and is sufficient to improve health, is limited. The Department for Environment Food and Rural Affairs website provides information and forecast on UK air quality, while the British Lung Foundation provides information on various measures to take according to air pollution levels (table 4).

Air cleaning systems are available commercially claiming to reduce indoor pollution—these can either remove particles and gaseous pollutants or have ultraviolet light technology to destroy indoor pollutants. All have their limitations, for example, large particles tend to settle before reaching filters, while gaseous pollutant filters may have short lifespans. These systems also use electricity—which may not be from sustainable sources. Improvement in our air quality will benefit the whole population with lasting health and economic advantages. We should aim to build cities in order to promote and improve the health of the population.

In conclusion, in the UK, the phasing out of the current diesel car, van and taxi fleet, and replacing this fleet with greener alternatives must be a pillar of exposure-reduction strategy. Changes that would support such an initiative are (1) more active travel supported by better public transport infrastructure, (2) providing electric charging points on residential streets, and (3) providing clinicians with the tools to discuss personal exposure reduction strategies with their patients.
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