Azo Dye Biodecolorization Enhanced by *Echinodontium taxodii* Cultured with Lignin

Yuling Han, Lili Shi, Jing Meng, Hongbo Yu, Xiaoyu Zhang

College of Life Science and Technology, Huazhong University of Science and Technology, Wuhan, China

Abstract

Lignocellulose facilitates the fungal oxidization of recalcitrant organic pollutants through the extracellular ligninolytic enzymes induced by lignin in wood or other plant tissues. However, available information on this phenomenon is insufficient. Free radical chain reactions during lignin metabolism are important in xenobiotic removal. Thus, the effect of lignin on azo dye decolorization in vivo by *Echinodontium taxodii* was evaluated. In the presence of lignin, optimum decolorization percentages for Remazol Brilliant Violet 5R, Direct Red 5B, Direct Black 38, and Direct Black 22 were 91.75% (control, 65.96%), 76.89% (control, 43.78%), 43.44% (control, 17.02%), and 44.75% (control, 12.16%), respectively, in the submerged cultures. Laccase was the most important enzyme during biodecolorization. Aside from the stimulating of laccase activity, lignin might be degraded by *E. taxodii*, and then these degraded low-molecular-weight metabolites could act as redox mediators promoting decolorization of azo dyes. The relationship between laccase and lignin degradation was investigated through decolorization tests in vitro with purified enzyme and dozens of aromatics, which can be derivatives of lignin and can function as laccase mediators or inducers. Dyes were decolorized at triple or even higher rates in certain laccase–aromatic systems at chemical concentrations as low as 10 μM.

Introduction

More than 700,000 tonnes of dyes are produced annually [1]. Azo, anthraquinone, sulfur, indigoid, triphenylmethyl (trityl), and phthalocyanine are most commonly used in industries, but azo derivatives account for approximately 70% of all synthetic colorants [2]. Processing and daily washing release up to 15% of the used dyes into the aquatic environment [3]. This practice adversely affects the aesthetic value of water bodies and poisons aquatic and terrestrial organisms [4]. All synthetic dyestuffs have chemical structures resistant to light, water, chemicals (e.g., oxidizing and reducing agents), and biological corrosion, making effluents from these compounds refractory pollutants that notably affect conventional biological wastewater treatments [5,6]. Traditional biological treatments of synthetic dyes such as activated sludge and biofilm are inexpensive and green alternatives. However, the efficiency of these techniques in the treatment of textile effluents considerably varies. Thus, a more economical and effective approach to dispose of wastewater with an extensive variety of dyes should be explored and implemented.

White-rot fungi (WRF) have attracted increasing scientific attention in environmental pollutant removal because of their unique ligninolytic enzyme system. Those enzymes are mainly extracellular and exhibit broad substrate specificity to decompose the complex and random phenylpropanoids polymeric structure of lignin [7]. Bioprocessing of dyes using certain WRF strains cultured under optimal fermentation conditions programmed for higher ligninolytic enzyme production have been investigated. Given the importance of extracellular enzymes in biodegradation, decolorization using crude or pure enzymes has been extensively explored. However, whole cells are superior to crude enzymes and much better than purified enzymes in generating nontoxic, harmless, or even completely mineralized end products; thus, dye decolorization significantly depends on an integrated fungal working system [8]. Hence, whole-cell treatment systems are further investigated for better utilization. Stains were found to fade out more rapidly in WRF cultures with low-cost lignocellulose substrates, comprising a wide variety of wastes from agricultural, forest, and food industries. Lignin-degrading enzymes were generally observed to be induced by natural additives than by synthetic media [9–13]. However, the mechanism behind the simultaneous increase in decolorization and enzyme production in natural substrates has not been elucidated. The biodegradation enhancement has been perceived to have resulted from the added oxidoreductases caused by the complex carbon resources or from the phenolic extracts from lignocellulose. Ligninolytic enzymes are generally promoted by nutrient structure (especially nitrogen limitation) regardless of the existence of a pollutant [8]. Moreover,
Materials and Methods

Ethics statement

No specific permissions were required for Shen-nung-chia Nature Reserve (Hubei Province, P.R. China) and the location was a scenic spot. The field studies did not involve endangered or protected species and the fungus was common in this location. Echinodontium taxodii 2538 was reported in our previous studies [22–25] and the GenBank accession number of this fungus was EF422215.

Microorganism and growth conditions

Echinodontium taxodii (GenBank accession number, EF422215) was isolated from rotten wood in Shen-nung-chia Nature Reserve (a scenic spot in Hubei Province, P.R. China) and preserved by our laboratory [24]. This strain was chosen for the present study because of the following reason. In our previous researches on biological pretreatment of lignocelluloses for energy production, the fungus exhibited great selective lignin-degrading ability [22–25]. Besides, E. taxodii was able to decolorize various dyes in a pretest of environment pollutants removal. The organism was maintained on maize straw plate at 4°C with periodic transfer and activated on basic medium agar plate at 28°C for 7 days [26].

Dyestuffs and chemicals

Remazol Brilliant Violet 5R (RBV5), Direct Red 5B (DR5B), Direct Black 36 (DB36) and Direct Black 22 (DB22) were purchased from Aladdin. 2,2'-Azinobis-3-ethylbenzothiazoline-6-sulfonic acid (ABTS) and lignin (CAS Number: 8068-05-1, Product Number: 370959) were obtained from Sigma-Aldrich. Veratryl alcohol and other fine chemicals were bought from Sinopharm Chemical Reagent Co. Ltd. China.

Effect of lignin on azo dye decolorization by Echinodontium taxodii

Three agar plugs with 1 cm diameter, which were punched from the periphery of 7-day agar plates, were cultivated in a stationary 250-mL flask containing 30 mL medium at 28°C. Quadruplicate flasks were used for each group. The media components of the four groups for each dye set in this study are shown in Table 1. Dye groups were used as the control for lignin–dye groups. Media containing only lignin were set to examine the effect of lignin on E. taxodii, which was inoculated in basic media as the corresponding control.

The basic medium per liter distilled water) contained 10.0 g glucose, 0.2 g ammonium tartrate, 2 g KH2PO4, 0.71 g MgSO4, 0.1 g CaCl2, and 20 mL trace element solution. The trace element solution had the following composition (per liter distilled water): 1 g NaCl, 0.5 g MnSO4·H2O, 0.1 g CuCl2, 0.1 g FeSO4·7H2O, 0.1 g ZnSO4·7H2O, 0.1 g CuSO4·5H2O, 0.01 g Na2MoO4·2H2O, and 0.01 g KAl(SO4)2·12H2O. The media were prepared as 1.2-fold-concentrated stock solutions and 25 mL of the media were split into 250-mL flasks. Azo dyes (RBV5, DR5B, DB36 and DB22, additional information on these dyes are shown in Table 2) were dissolved in distilled water with a final concentration of 600 mg/L. Then, distilled water, the media and dyes were autoclaved at 115°C for 20 min. After that, dye solutions or distilled water (5 mL) were added to each flask sterile. Different concentrations of lignin (0.1, 0.3, and 0.5 g/L) were sterilized with ultraviolet light for 30 min before being added to the sterilized media. Considering the disoriented state of watersoluble components, lignin was washed thoroughly with distilled water and then freeze-dried to constant weight before use. The cultures of each group in quadruplicate flasks were taken on the day after inoculation. The fermentation liquor was obtained by filtering out the mixture of mycelia and lignin residues before centrifugation at 10,000×g for 10 min to remove water-insoluble lignin. The mycelia were oven-dried to a constant weight and were calculated as the biomass after the solid mixture was washed through a 100-mesh filter gauze to remove the lignin and media residues.

In order to evaluate the adsorption by lignin, different concentrations of lignin were added to media containing dyes sterilely, without inoculation. The samples were taken at the same time with cultures mentioned above, and centrifuged at 10,000×g for 10 min to remove water-insoluble lignin. The liquid supernatant was used to determine the decolorization of lignin adsorption.

Decolorization was monitored by measuring the absorbance of the fermentation broth at the maximum wavelength using a spectrophotometer. Decolorization percentage (%) was calculated as follows: decolorization percentage (%) = \( \frac{[A_0 - A]}{A_0} \times 100\).
Table 1. The media used in this study.

| contents     | control | lignin | dye | Lignin-dye |
|--------------|---------|--------|-----|------------|
| basic medium | +       | +      | -   | +          |
| lignin       | -       | -      | +   | -          |
| dye          | -       | -      | -   | +          |

+ With the content listed on the left added
- Without the content listed on the left added

doi:10.1371/journal.pone.0109786.t001

where \( A_0 \) is the initial absorbance of the dye, and \( A \) is the absorbance of dye with time.

Laccase activity was determined by measuring the increase in \( A_{420} \) with 0.5 mM ABTS as the substrate in 50 mM sodium acetate buffer, pH 4.5 (\( \varepsilon_{420} = 29,300 \) M\(^{-1}\)cm\(^{-1}\)) [27]. Manganese peroxidase (MnP) activity was assayed using the method of Warish, Valli, Gold [28] and by measuring the increase in absorbance at 270 nm with 1 mM Mn\(^2+\) as the substrate in 50 mM malonic acid buffer, pH 4.5 (\( \varepsilon_{270} = 11,590 \) M\(^{-1}\)cm\(^{-1}\)). The reaction was started by adding 0.4 mM H\(_2\)O\(_2\). Lignin peroxidase (LiP) activity was determined using the method of Tien, Kirk [29], with veratryl alcohol as a substrate in 0.1 M sodium tarteit, pH 3.0 (\( \varepsilon_{310} = 9,300 \) M\(^{-1}\)cm\(^{-1}\)). All reactions were performed at room temperature.

Effect of lignin-related compounds on azo dye decolorization by purified laccase of _E. taxodii_

Laccase was the only ligninolytic enzyme secreted by this strain during the whole dye removal period, and lignin might affect biodecolorization through this key oxidase. Laccase was purified by salting out with (NH\(_4\))\(_2\)SO\(_4\), hydrophobic interaction chromatography with Phenyl Sepharose 6 Fast Flow, and ion exchange chromatography with diethylethanolamine (DEAE)–sepharose Fast Flow.

The _E. taxodii_ culture medium was passed through a filter paper to remove fungal mycelia. The supernatant was loaded onto a Phenyl Sepharose 6 Fast Flow column (1.6 x 20 cm) equilibrated with the same buffer. Unbound proteins were removed by washing the column at a flow rate of 1 mL·min\(^{-1}\). Decreasing concentrations of (NH\(_4\))\(_2\)SO\(_4\) (0.8 M to 0 M) were used to elute the bound proteins, and the collected fractions (5 mL each) were subjected to enzymatic assay and protein measurement (\( A_{280} \)). The active fractions were pooled, dialyzed against 20 mM sodium acetate–acetic acid buffer (pH 6) overnight, and concentrated by PEG 20000. The concentrate was then applied to a DEAE–Sepharose Fast Flow column (1.6 x 20 cm) pre-equilibrated with the same buffer. Proteins were eluted with a linear gradient of NaCl (0 M to 0.8 M) at a flow rate of 1.5 mL·min\(^{-1}\). Subsequently, fractions containing laccase activity were pooled, concentrated, and desalted. Enzyme purity was then confirmed by sodium dodecyl sulfate–polyacrylamide gel electrophoresis (SDS-PAGE).

Dyes were decolorized in a test tube using purified laccase at room temperature. The reaction mixture (5 mL) contained sodium acetate–acetic acid buffer (50 mM, pH 4.8), purified laccase (final concentration of 1 U/mL), and dyes (final concentration of 100 mg/L) in the presence or absence of lignin-related phenols. The reaction was initiated by adding the enzyme solution under mild shaking conditions. The duration of decolorization was determined by measuring the absorbance at the maximum wavelengths with a Mapada™ UV-1600 spectrophotometer. Decolorization was calculated as described previously. A control test containing the same amount of heat-denatured laccase was conducted in parallel. All reactions were performed in triplicate.

To evaluate dye decolorization using different natural mediators, 12 lignin-related phenols as natural mediators were applied in the decolorization of dyes. These natural mediators (final concentration of 100 mg/L) in the presence or absence of lignin-related phenols. The reaction was initiated by adding the enzyme solution under mild shaking conditions. In order to evaluate the potential interference of the 12 lignin-related phenols as natural mediators on decolorization measurement, another control test without laccase and dyes were set. The results showed that these phenols had no absorption at those wave lengths, so decolorization was calculated as described previously. The decolorization of dyes was calculated after reaction for 2 h. Then, lower concentrations (10, 50, 100, 250 and 500 μM) of the efficient mediators were applied to the buffer under the same conditions mentioned above for further investigation of the

Table 2. Characteristics of the synthetic dyes studied in this work.

| Dyes   | CAS      | Molecular Formula | \( \lambda_{max}(\text{nm}) \) | Number of azo bonds |
|--------|----------|-------------------|-----------------------------|---------------------|
| RBV5   | 12226-38-9 | C\(_{20}\)H\(_{16}\)N\(_{2}\)Na\(_{3}\)O\(_{15}\)S\(_{4}\) | 560 | 1 |
| DR5B   | 2610-11-9 | C\(_{30}\)H\(_{16}\)N\(_{2}\)Na\(_{2}\)O\(_{6}\)S\(_{2}\) | 510 | 2 |
| DB38   | 1937-37-7 | C\(_{34}\)H\(_{20}\)N\(_{2}\)Na\(_{2}\)O\(_{8}\)S\(_{2}\) | 550 | 3 |
| DB22   | 6473-13-8 | C\(_{46}\)H\(_{20}\)N\(_{2}\)Na\(_{2}\)O\(_{15}\)S\(_{4}\) | 482 | 4 |

doi:10.1371/journal.pone.0109786.t002
efficiency of these mediators. The decolorization of dyes was detected at different time intervals (from 5 min to 240 min).

**Results and Discussion**

**Effect of lignin on dye decolorization by *E. taxodii***

Azo reactive dyes RBV5, DR5B, DB38, and DB22 were chosen because they are extensively used in the textile industry worldwide. These azo dyes with 1 to 4 azo groups were tested to establish the relationship between the decolorization effect of *E. taxodii* (with/without lignin) and the structures of azo dyes. *E. taxodii* could remove all four dyes (for as long as 2 wk; data not shown), and the total decolorization percentages of the RBV5, DR5B, DB38, and DB22 controls on day 6 were 65.96%, 43.78%, 53.02%, and 57.20%, respectively. The results showed that the decolorization effect differed from one dye to another, and there was no correlation between the difficulty degree of decolorization and the number of azo bonds. However, compared with the controls, lignin addition enhanced the biological treatment process, no matter how many azo linkages there was (Fig. 1).

In the case of RBV5, *E. taxodii* performed fairly in the first few days. Maximal differences in the decolorization percentage were observed between the control (65.96%) and lignin tests (75.54%, 100 mg/L; 80.64%, 300 mg/L; and 91.75%, 500 mg/L) on day 6, which were 15%, 22%, and 39% higher than that in non-lignin media. On day 7, lignin-free cultures still had dye residues at approximately 25 mg/L, whereas all the test groups had less than 10 mg/L residue. For DR5B, lignin resulted in a notable increase in the decolorization percentage from day 3, and the discrepancies were 5.69% (100 mg/L), 11.95% (300 mg/L), and 7.47% (500 mg/L). On the last day (day 7), control media still showed lower decolorization (75.99%) than the test groups by 16.29% (100 mg/L), 13.37% (300 mg/L), and 18.55% (500 mg/L). In addition, the discrepancies in the decolorization percentage first increased and then decreased along the degradation period. The maximal differences in the decolorization percentage between control and test groups varied, as follows: 100 mg/L, 16.77% on day 6; 300 mg/L, 31.67% on day 5; and 500 mg/L, 33.11% on day 6. For RBV5 and DR5B, the enhancement in decolorization seemed to have intensified with an increase in lignin dosage.

*E. taxodii* in lignin-containing media initially decolorized dye DB38 quickly and then the activity slowed down. However, the fungus worked at nearly constant speed in the blank culture. Different concentrations of lignin did not result in significant differences in dye degradation, but remarkable enhancements were observed when compared with lignin-free experiments in the first few days. Maximum differences in dye decolorization percentage compared with the blank were observed on day 2 or

**Figure 1. Effect of lignin on the decolorization of different dyes by *E. taxodii***. (A): RBV5. (B): DR5B. (C): DB38. (D): DB22. Dyes concentration: 100 mg/L. Lignin contents: ■, 0 mg/L; ●, 100 mg/L; ▲, 300 mg/L; ▼, 500 mg/L.
doi:10.1371/journal.pone.0109786.g001
3 after inoculation, as follows: 26.42% (100 mg/L on day 3), 23.74% (300 mg/L on day 3), and 25.67% (500 mg/L on day 2). Differences gradually decreased, and the control even performed fairly on day 6 (Fig. 1). Differences in decolorization curves caused by lignin indicated that *E. taxodii* treatment of DB38 may follow disparate pathways at first. Similar results were obtained when the fungus reacted with DB22, and more significant promotion than DB38 degradation was recorded. From days 1 to 6, the dye was eliminated much more rapidly along with lignin than the control, and the differences reached maximum values on day 3, as follows: 25.92% (100 mg/L), 31.81% (300 mg/L), and 32.59% (500 mg/L). Afterward, no significant differences were detected.

Dye bioadsorption using dead or living biomass of WRF has been extensively reviewed in several recent papers [30,31]. Hence, the absorption function of biomass and lignin was considered in determining the relationship between biological degradation and lignin metabolism. Stationary surface culture was adopted in this study. Results showed that lignin had different absorption patterns on the four dyes. Peak values were all reached on day 2 at the highest lignin concentration with less than 10% absorption as

![Figure 2. Effect of lignin on biomass during the decolorization of dyes by *E. taxodii*.](image)

**Table 3.** Effect of lignin on the adsorptions of azo dyes.

| Adsorption Percentage of Azo dyes (%) | Lignin content (mg/L) |
|--------------------------------------|-----------------------|
|                                       | 100       | 300       | 500       |
| RBV5                                 | 3.72(0.62) | 4.31(0.53) | 7.21(1.62) |
| DRSB                                 | 3.91(0.78) | 6.75(2.09) | 9.09(1.02) |
| DB38                                 | None      | None      | None      |
| DB22                                 | 3.62(0.78) | 5.76(1.36) | 8.96(1.02) |

a Average Percentage with Standard Deviations in Parentheses; N = 3.
shown in Table 3. In addition, RBV5 and DR5B showed no more than 10% adsorptive decolorization, whereas the actual dye removal was only approximately at 10% independent of the presence of lignin. We observed that lignin was not an ordinary adsorbent for RBV5 and DR5B when cooperating with living fungus. For DB38 and DB22, decolorization differences between lignin–dye groups and the dye controls (>20%) were larger than the highest adsorption values (<10%) on day 2 or 3 (Fig. 1). This result indicated that lignin might enhance dye elimination through some mechanism other than physical–chemistry adhesion. By contrast, biomass from culture inoculated with lignin did not grow faster than that from the control group. However, decolorization seemed to significantly speed up in the former, indicating the absence of difference in the biomass activity on absorption (Fig. 2). This result confirmed that enhanced decolorization was influenced by the metabolism of added lignin.

Effect of lignin on ligninolytic enzymes during the decolorization by *E. taxodii*

Fungi degrade lignin or azo dyes through oxidation, probably involving enzymes such as LiP, MnP, and laccase. Meanwhile, toxic aromatic compounds and lignin (lignosulfonate and synthetic lignin) were shown to stimulate the production of laccase, LiP, and MnP [32]. However, in our research, among the main ligninolytic enzymes, only laccase was found in the liquid culture of *E. taxodii* throughout the biological decolorization (Fig. 3). Once lignin was present, significantly higher values of laccase activity were detected in the culture media, even when only lignin was present in the substrate with no dyes. Enzymatic activities were approximately twice to thrice higher than that in the lignin-dye-free blank (<5 IU/L). In addition, dyes alone can stimulate the yield of laccase which generally increased to different extents with the extension of incubation time, as follows: RBV5, 77.95 IU/L; DR5B, 231.92 IU/L; DB38, 53.03 IU/L; and DB22, 48.13 IU/L on the last day of degradation. The stimulation effects of dyes on laccase production during cultivation, which could be regarded as xenobiotic response to detoxification, have also been reported [33]. However, the fungus incubated with lignin produced more laccase and enhanced dye removal (Fig. 3). In the case of RBV5 and DR5B (Figs. 1A, 1B, 3A, and 3B), decolorization percentages were initially not enhanced along with the increase in laccase. Afterward, dyes fade out at higher rates when lignin was added in increasing amounts. However, laccase activities did not occur in similar patterns, i.e., no positive correlation was observed between the two phenomena. For DB38 and DB22 (Figs. 1C, 1D, 3C, and 3D), the decolorization curves were poorly consistent with enzymatic activity detection. All results showed that dyes decolorized in the presence of lignin were affected not only by laccase but also by lignin metabolism. Little information

Figure 3. Effect of lignin on laccase activities during the decolorization of dyes by *E. taxodii*. (A): RBV5. (B): DR5B. (C): DB38. (D): DB22. Dyes concentration: 100 mg/L. Lignin contents: ■, 0 mg/L; ●, 100 mg/L; ▲, 300 mg/L; ◀, 500 mg/L. doi:10.1371/journal.pone.0109786.g003
on the combined degradation of pollutants and lignin or the resolution of the inner mechanism is available. According to Tychanowicz, Zilly, de Souza, Peralta [34], structurally different dyes were decolorized because of the high level of laccase generated in response to the presence of soluble phenolic compounds in the glucose/ammonium tartrate–corn cob solid-state medium. The fungi produced laccase and removed color at the same time with the supplementation of lignocellulose, providing an alternative approach to deal with textile wastewater and increasing the economic value added by producing laccase [9]. The underlying promoting mechanism should be uncovered to provide a theoretical guideline with strong feasibility in depleting diverse dyes or pollutants.

Thus, Figs. 1 to 3 show that lignin can influence dye removal through laccase, the only oxidase during biotreatment, in some pathway different from the common improvement of enzymatic activity. According to papers published worldwide, WRF secrete a large amount of volatile aromatic compounds, giving rise to natural fragrances, providing a promising source of flavors and aromas. Gutierrez, Caramelo, Prieto, Martinez, Martinez [35] reported that simple aromatic compounds could be generated by fungi with or even without aromatic origins. However, the products varied in quantities and kinds as stimulators were added. Low-molecular-weight aromatic acids have been identified during the long-term degradation of lignin in spruce wood by \textit{Phanerochaete chrysosporium} by high-performance liquid chromatography and after acetylation and methylation by gas chromatography/mass spectrometry (GC/MS) [19]. Additionally, some of these chemicals were low-molecular-weight phenol derivatives and considered potential laccase substrates, which could also work as natural mediators for the oxidation reactions by laccase. Johannes, Majcherczyk [36] tested different kinds of mediators for laccase in

![Figure 4. SDS-PAGE analyses of purified laccase of \textit{E. taxodii}.](image)

Lane 1, Molecular weight marker; Lane 2 and 3, purified laccase.
doi:10.1371/journal.pone.0109786.g004

![Figure 5. Evaluating for nature mediators based on decolorization of DR5B, DB22 and DB38 (100 mg/L) in presence of mediator (1 mM) for 2 h.](image)

doi:10.1371/journal.pone.0109786.g005
the oxidation of polycyclic aromatic hydrocarbons (PAH) and determined that natural compounds such as phenol, aniline, 4-hydroxybenzoic acid, and 4-hydroxybenzyl alcohol are efficient as synthetic chemicals ABTS and 1-hydroxybenzotriazole. Moreover, \( p \)-hydroxybenzoinic acids, namely, \( p \)-coumaric (4-hydroxybenzoinic) acid, ferulic (3-methoxy-4-hydroxybenzoinic) acid, and sinapic (3,5-dimethoxy-4-hydroxybenzoinic) acid, were used as natural mediators for laccase oxidation of recalcitrant dyes and PAH [37]. Thus, we infer that the metabolites of lignin may act as redox mediators in the oxidation of dyes by laccase, the main oxidative enzyme in decolorization.

In this study, lignin metabolite detection was also conducted. However, we did not observe similar results on the chromatogram map from GC/MS (data not shown) [19]. We infer that lignin degradation in large quantities started during the secondary metabolism of WRF and was a long-term process. Thus, low-molecular-weight lignin derivatives were likely in trace amount and were difficult to identify during the short-term decolorization. Moreover, the metabolic intermediates were probably initiated by the increase in laccase during the active growth phase (based on biomass accumulation) when the fungus did not function on lignin degradation. Hence, the metabolic intermediates were difficult to accumulate and often existed briefly. In addition, laccase can graft aromatic substrates back onto lignin [38,39]. Thus, various lignin-derived aromatics were studied in vitro in mediating degradation by purified laccase.

**Effect of lignin related compounds on the azo dyes decolorization by purified laccase of *E. taxodii***

Laccase is the most potent oxidase involved in the degradation of lignin and dyes. Laccase was purified from *E. taxodii* cultures to homogeneity by salt fractionation with ammonium sulfate, hydrophobic interaction, and ion exchange chromatography. Approximately 23.3-fold purification was achieved, with an overall yield of 38.8% (data not shown). The purified laccase showed a single protein band on SDS-PAGE (Fig. 4).

The purified laccase of *E. taxodii* slowly decolorized RBV5, DR5B, DB22, and DB38 in the absence of mediators, with corresponding decolorization percentages of 1%, 2%, 10%, and 7% after 2 h reaction. However, the dye decolorization percentages by laccase were enhanced in the presence of natural mediators, except that of RBV5, which did not disappear more rapidly with mediators present. Among the 13 aromatics (phenolic aldehydes, ketones, and acids), acetoxyxynine and syringaldehyde significantly promoted the decolorization of DR5B, DB22, and DB38. Decolorization of the last two dyes was also accelerated by \( p \)-coumaric acid, acetovanillone, and vanillin (Fig. 5). However, we did not obtain any suitable redox mediators in the candidates for RBV5. Among the most efficient mediators, syringaldehyde provided the highest decolorization percentages for DR5B and DB38. The decolorization percentages increased 37-fold and 3.7-fold, respectively, from 2% to 74% for DR5B and from 7% to 51% for DB38. Compared with syringaldehyde, \( p \)-coumaric acid was more effective in the decolorization of DB22. The decolorization percentage increased 4.5-fold from 10% to 45%. Camarero, Ibarra, Martínez, Martínez [40] determined that syringaldehyde, acetoxyxynine, vanillin, acetovanillone, methyl vanillate, and \( p \)-coumaric acid, which were ubiquitous laccase producers of basidiomycetes in decayed wood and forest soils, significantly promoted the oxidation of recalcitrant dyes by laccase.

The influences of natural mediator concentrations on the decolorization of DR3B, DB22, and DB38 (4 h treatment) are shown in Fig. 6. The decolorization percentages of the three dyes were correlated with the concentration of mediators. Laccase in the presence of \( p \)-coumaric acid produced a high decolorization percentage for DB22, whereas the presence of syringaldehyde promoted the decolorization percentage for DB38 at a low concentration (10 \( \mu \)M). By contrast, the decolorization percentage for DR5B was directly proportional to syringaldehyde concentration. The addition of syringaldehyde of 1 mM concentration enhanced the decolorization percentage from 0.2% to 74.0% after reaction for 30 min, whereas the decolorization percentage did not change at 10 \( \mu \)M concentration of syringaldehyde.

Although different mediators apparently influence the efficiency of laccase to decolorize the dyes in vitro [40], actual decolorization by *E. taxodii* in the presence of lignin might be affected by the various previously tested lignin-derived compounds. Meanwhile, the contents of each chemical vary from those obtained in this study and were possibly much lower. The result showed that aromatics like \( p \)-coumaric acid and syringaldehyde enhanced decolorization by purified laccase, and the enhancement was positively related with the concentration of aromatics. When phenols at a low concentration, they did not increase the decolorization as effectively as they were at a concentration of 1 mM. It was presumed that lignin decomposition product a small quantity of aromatics, and they slowly boosted decolorization in the cultures. By contrast, Chen, Chang, Kirk [19] studied the degradation products of lignin in decayed wood. Monomers and dimers were detected, in which dimers, e.g., dihydroxydianillic acid and 2′-hydroxy-2,3′-dimethoxydiphenylether, were also produced from monomer polymerization by laccase [41]. Thus,
dimers could be detected when aromatics mediated lignin or dye oxidation by laccase. However, natural mediators in the fungal system are not always at a detectable concentration (they are difficult to accumulate and often existed only briefly, but usually performing better than synthetic ones), and the derivatives are in low quantity to be identified, contrary to that in vitro which is set to a much higher level in lignin or xenobiotic compound degradation. Thus, *E. taxodii* degraded lignin and yielded hundreds of metabolites, among which some chemicals worked as redox mediators in laccase oxidation of dyes.

**Conclusions**

In the current work *E. taxodii* was observed to degrade four azo dyes more efficiently when lignin was present in the media. Laccase, the dominant enzyme in the biodegradation, was found to be induced by lignin. Experiments in vitro with purified laccase demonstrated that contribution of lignin-related aromatics as laccase mediators might be the explanation for the improvement in decolorization.

**Supporting Information**

**File S1** Data of Figure 1. Effect of lignin on the decolorization of different dyes by *E. taxodii*. Table S1, RBV5. Table S2, DR5B. Table S3, DB38. Table S4, DB22. (XLS)

**References**

1. Burkinshaw S, Salihu G (2013) The wash-off of dyesings using interstitial water Part 4: disperse and reactive dyes on polyester/cotton fabric. Dyes Pigment 99: 540–560.
2. Erkurt EA, Erkurt HA, Uneyar A (2010) Decolorization of azo dyes by white rot fungi. Biodegradation of Azo Dyes. Springer. pp. 157–167.
3. Passetti G, Telfe A, Kalyani D, Govindbhai S (2010) Decolorization and detoxification of sulfonated azo dye methyl orange by *Kocuria rosea* MTCC 1532. J Hazard Mater 176: 503–509.
4. Upton de Souza SMAG, Morgarini E, Upton de Souza AA (2007) Toxicity of textile dyes and their degradation by the enzyme horseradish peroxidase (HRP). J Hazard Mater 147: 1073–1078.
5. Isik M, Sponek EF (2007) Fate and toxicity of azo dye metabolites under batch long-term anaerobic incubations. Enzyme Microb Technol 40: 934–939.
6. de Souza SMdAGU, Bonilla KAS, de Souza AAU (2010) Removal of COD and color from hydrolyzed textile azo dye by combined ozonation and biological treatment. J Hazard Mater 179: 35–42.
7. Gao D, Du L, Yang J, Wu W-M, Liang H (2010) A critical review of the application of white rot fungus to environmental pollution control. Crit Rev Biotechnol 30: 79–77.
8. Kowshik P, Malik A (2009) Fungal dye decolorization: recent advances and future potential. Environ Int 35: 127–141.
9. Liu J, Cai Y, Liao X, Huang Q, Hao Z, et al. (2012) Simultaneous Laccase Production and Color Removal by Culturing Fungus *Pycnoporus coccineus* sp. SYBG-L3 in a Textile Wastewater Effluent Supplemented with a Lignocellulosic Waste *Phragmites australis*. Bull Environ Contam Toxicol 89: 269–273.
10. Ozmen N, Yesilada O (2012) Valorization and biodecolorization of dye adsorbed on lignocelluloses using white rot fungi. BioResources 7: 1656–1665.
11. Keliang Y, Wang H, Zhang X, Yu H (2009) The effect of biological degradation of reduction of azo dye by *Pleurotus ostreatus* under solid-state cultivation. J Microbiol Biotechnol 19: 1421–1430.
12. Yan K, Wang H, Zhang X (2009) Biodegradation of Crystal Violet by low molecular mass fraction secreted by fungus. J Biocatal Biofuel 4: 121–424.
13. Verma P, Madamwar D (2002) Production of ligninolytic enzymes for dye decolorization by cocultivation of white-rot fungi *Phanerochaete chrysosporium* and *Sphingomonas* involved in anaerobic reduction of azo dyes by *Sphingomonas* sp. Strain BN6 and effect of artificial redox mediators on the rate of azo dye reduction. Appl Environ Microbiol 63: 3691–3694.
14. Yu H, Zhang X, Song L, Liu J, Xu C, et al. (2010) Evaluation of white-rot fungus-assisted alkaline/oxidative pretreatment of corn straw undergoing enzymatic hydrolysis by cellulase. J Biosci Bioeng 110: 660–664.
15. Ma F, Yang N, Xu C, Yu H, Wu J, et al. (2010) Combination of biological pretreatment with mild acid pretreatment for enzymatic hydrolysis and ethanol production from water hyacinth. Bioresource Technol 101: 9600–9604.
16. Zhang X, Yu H, Huang H, Liu Y (2007) Evaluation of biological pretreatment with white rot fungi for the enzymatic hydrolysis of bamboo culms. Int Biodeterior Biodegrad 60: 139–164.
17. Bugg TD, Alamad MA, Hardiman EM, Rahmanpour R (2011) Pathways for degradation of lignin in bacteria and fungi. Nat Prod Rep 28: 1883–1896.
18. Aupper M, Bharti HS, Ashraf M, Legge RL (2008) Recent developments in biodegradation of industrial pollutants by white rot fungi and their enzyme system. Biodegradation 19: 771–783.
19. Chen C-L, Chang H-M, Kirk TK (1985) Carboxylic acids produced through oxidative cleavage of aromatic rings during degradation of lignin in spruce wood by *Phanerochaete chrysosporium*. J Wood Chem Technol 3: 43–57.
20. Piccolo A (2002) The supramolecular structure of humic substances: a novel understanding of humus chemistry and implications in soil science. Adv Agron. pp. 57–134.
21. Kudlich M, Keck A, Klein J, Stoka A (1997) Localization of the enzyme system involved in anaerobic reduction of azo dyes by *Sphingomonas* sp. Strain BN6 and effect of artificial redox mediators on the rate of azo dye reduction. Appl Environ Microbiol 63: 3691–3694.
22. Yu H, Zhang X, Song L, Liu J, Xu C, et al. (2010) Evaluation of white-rot fungus-assisted alkaline/oxidative pretreatment of corn straw undergoing enzymatic hydrolysis by cellulase. J Biosci Bioeng 110: 660–664.
23. Ma F, Yang N, Xu C, Yu H, Wu J, et al. (2010) Combination of biological pretreatment with mild acid pretreatment for enzymatic hydrolysis and ethanol production from water hyacinth. Bioresource Technol 101: 9600–9604.
24. Zhang X, Yu H, Huang H, Liu Y (2007) Evaluation of biological pretreatment with white rot fungi for the enzymatic hydrolysis of bamboo culms. Int Biodeterior Biodegrad 60: 139–164.
25. Yu H, Guo G, Zhang X, Yan K, Xu C (2009) The effect of biological pretreatment with the selective white-rot fungus *Echinodontium taxodii* on enzymatic hydrolysis of softwoods and hardwoods. Bioresour Technol 100: 3170–3173.
26. Yang N, Zeng Y, Ma F, Zhang X, Yu H (2010) Effect of biopretreatment on thermogravimetric and chemical characteristics of corn stover by different white-rot fungi. Bioresource Technol 101: 5475–5479.
27. Wolfenden BS, Wilham RL (1982) Radical-cations as reference chromogens in kinetic studies of one-electron transfer reactions: pulse radiolysis studies of 2,2'-azinobis(3-ethylbenzothiazoline-6-sulfonate). J Chem Soc Perkin Trans 2: 805–812.
28. Warishi H, Valli K, Gold MH (1992) Manganese (II) oxidation by manganese peroxidase from the basidiomycete *Phanerochaete chrysosporum*. Kinetic mechanism and role of chelators. J Biol Chem 267: 23681–23685.
29. Tien M, Kirk TK (1988) Lignin peroxidase of *Phanerochaete chrysosporum*. Methods Enzymol 161: 238–249.
30. Binupriya A, Sathishkumar M, Swaminathan K, Kuz C, Yun S (2008) Comparative studies on removal of Congo red by native and modified mycelial pellets of *Trametes versicolor* in various reactor modes. Bioresource Technol 99: 1080–1088.
31. Lin Y, He X, Han G, Tian Q, Hu W (2011) Removal of Crystal Violet from aqueous solution using powdered mycelial biomass of Ceriporia lacera. J Environ Sci 23: 2055–2062.

32. Rogalski J, Lundell TK, Leonowicz A, Hatakka AI (1991) Influence of aromatic compounds and lignin on production of ligninolytic enzymes by Phlebia radiata. Phytochemistry 30: 2869–2872.

33. Puvaneswari N, Muthurkishnan J, Gunasekaran P (2006) Toxicity assessment and microbial degradation of azo dyes. Indian J Exp Biol 44: 618.

34. Tychanowicz GK, Zilly A, de Souza CGM, Peralta RM (2004) Decolourisation of industrial dyes by solid-state cultures of Pleurotus pulmonarius. Process Biochem 39: 855–859.

35. Gutierrez A, Caramelo L, Prieto A, Martinez MJ, Martinez AT (1994) Anisaldehyde production and aryl-alcohol oxidase and dehydrogenase activities in ligninolytic fungi of the genus Pleurotus. Appl Environ Microbiol 60: 1783–1788.

36. Johannes C, Majcherczyk A (2000) Natural mediators in the oxidation of polycyclic aromatic hydrocarbons by laccase mediator systems. Appl Environ Microbiol 66: 524–528.

37. Camarero S, Canas AI, Nousiainen P, Record E, Lomascolo A, et al. (2008) p-Hydroxycinnamic acids as natural mediators for laccase oxidation of recalcitrant compounds. Environ Sci Technol 42: 6703–6709.

38. Moilanen U, Kellock M, Galkin S, Vilkari L (2011) The laccase-catalyzed modification of lignin for enzymatic hydrolysis. Enzyme Microb Technol 49: 492–498.

39. Rencoret J, Aracri E, Gutiérrez A, del Río JC, Torres AL, et al. (2014) Structural insights on laccase biografting of ferulic acid onto lignocellulosic fibres. Biochem Eng J 86: 16–23.

40. Camarero S, Ibarra D, Martinez MJ, Martinez AT (2005) Lignin-derived compounds as efficient laccase mediators for decolorization of different types of recalcitrant dyes. Appl Environ Microbiol 71: 1775–1784.

41. Ibrahim V, Volkova N, Pyo S-H, Mamo G, Hatti-Kaul R (2013) Laccase catalyzed modification of lignin subunits and coupling to p-aminobenzoic acid. J Mol Catal B: Enzym 97: 45–53.