Beyond Strict Regulations to Achieve Environmental and Economic Health—An Optimal PM$_{2.5}$ Mitigation Policy for Korea

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Abstract: Growing concern about particulate matter (PM$_{2.5}$) pressures Korea to reduce the health risks associated with its high dependency on fossil fuels. The Korean economy relies heavily on large thermal power plants—a major source of PM$_{2.5}$ emissions. Although air quality regulations can negatively impact local economies, the Korean government announced two strict air quality mitigation policies in 2019. We develop a regional static computable general equilibrium model to simulate the economic and environmental impacts of these policies under alternative hypothetical scenarios. We separate two regions, Chungcheongnam-do, the most polluted region, and the rest of the country, in our model. As policy options, we introduce a regional development tax and a tradable market for PM emission permits, similar to an air pollution tax and a carbon permits market, respectively. The results show that allowing higher tax rates and a tradable permits market gives the optimal combination, with the PM$_{2.5}$ emissions reduced by 2.35% without sacrificing economic growth. Since alternative options present, for example, a 0.04% loss of gross domestic product to reduce PM emissions by the same amount, our results here may present a new policy paradigm for managing air pollutants such as PM$_{2.5}$.

Keywords: air pollutants; particulate matter (PM$_{2.5}$); thermal power generation; auxiliary benefit; regional computable general equilibrium model

1. Introduction

Among the adverse effects of economic growth, air quality degradation driven by fossil fuel energy sources represents a mitigation target in most countries. Air pollutants from fossil fuel power generation such as sulfur dioxide (SOx) and nitrogen oxides (NOx) caused approximately 3.7 million deaths worldwide in 2012, the largest single environmental health risk [1]. However, most economic activities in many countries rely heavily on fossil fuel energy sources, making air pollution all but inevitable. In addition, ambient particulate matter such as fine particulate matter (PM$_{2.5}$) causes the most substantial health risks, such as cardiovascular and cerebrovascular diseases [2]. Epidemiological studies elaborate a robust relationship between long-term exposure to ambient air pollution and premature mortality. In an example going back decades, a dramatic increase in particulate matter in...
the 1952 “London Fog” greatly increased the infant mortality rate [3]. More recently, Apte et al. [4] addressed the mortality from air pollution and the potential health benefits of reducing ambient PM$_{2.5}$. Aggregate effects from an aggressive global program to reduce PM$_{2.5}$ implemented by the WHO (World Health Organization) could substantially reduce the mortality in more polluted regions, such as China and India. Since air pollutants are a transboundary problem, severe air quality degradation in China also significantly impacts the Korean peninsula [5]. Prevailing westerlies deliver air pollutants from China, exacerbating the air quality in the late-industrialized economy of the Republic of Korea (hereafter, “Korea”) [6]. Some of the literature indicates that there is a significant improvement in air quality in China recently [7,8], but air quality degradation is one of the most serious social problems that needs to be resolved in Korea.

Among the Organization for Economic Co-operation and Development (OECD) member countries, Korea has a notorious air quality level, with a mean population exposure to PM$_{2.5}$ of 25.1 µg/m$^3$ in 2017 [9]. In 2017, the Korean government launched “Comprehensive Measures for Particulate Matter Management” to lower the domestic PM$_{2.5}$ emissions by 35% by 2024 [10]. To implement this plan, detailed action plans suggest reducing the number of coal power generation plants and restricting the operation of high-emission vehicles. The plan emphasizes the importance of controlling both primary and secondary PM$_{2.5}$ emissions simultaneously. Thermal power plants primarily fuel late-industrialized economies like Korea, so using strict air pollutant mitigation policies to limit particulate emissions from these plants predictably drags on economic productivity. Much economic literature has attempted to predict how a reduction in air pollutants will correlate with specific policy interventions. Since a computable general equilibrium (CGE) model simulates a general equilibrium where utility is maximized through consumption and production activities with respect to major economic agents’ goals [11], it is widely used for analyzing the interaction between the production and consumption of energy-related goods and services, where ambient particulate matter is a secondary product of energy production, and where the impacts of mitigation policies can affect economic performance [12].

Nam et al. [12] showed how a CGE model can incorporate air pollutant reduction policies. Nam et al. compared a European “Clean Air Policy Package”, whose projected health benefits for the European Union more than offset the projected economic burden, to China’s need to reduce its coal dependence to meet its pollution reduction targets. PM pollution lost China 112 billion US dollars in 2005. Xiao et al. [13] examined the effects of an “air pollution tax” in China, utilizing a recursive dynamic CGE model to simulate the effects of policy options for mitigating air pollutants. The industry-specific effects were positive or negative depending on the volume of particulate matter emission. Similarly, Wu et al. [14] calculated the change in PM$_{2.5}$ emissions resulting from trade across provinces, and concluded that more than 30% of particulate matter was caused by multiregional economic activities. Wu et al. [14] concluded that the Chinese government needs to establish different air pollutant mitigating policies that consider heterogeneous needs across stakeholders. Korea has a similar desire to reduce air pollutants, including PM$_{2.5}$, due to its rapid economic growth fueled by a dramatic increase in energy consumption [15]. Recently, Oh et al. [6] developed a static CGE model to simulate the economic impact of air pollutant reduction policies and to test the interactions between those policies in Korea. The results show that the costs of providing clean air are high, up to 1.75% of the gross domestic product, and would cause an asymmetrical loss to high-emission over low-emission industries. To the best of the authors’ knowledge, Oh et al. [6] made the first attempt to assess the Korean air-pollutant reduction policy (i.e., “Comprehensive Measures for Particulate Matter Management”) using a CGE model. However, they assume that the government provides an emissions credit market that may be established in Korea. Xiao et al. [13] introduced an air pollution tax to high-emission industries that is similarly applied in Korea.

In this research, we aim to find a feasible solution for ambient particulate matter reduction polices using a model that allows a mix of policies, including the simultaneous application of an “air pollution tax” and an “emissions credit trading market.” A regional static CGE model is developed to assess the
PM$_{2.5}$ emissions from multiregional business activities in Korean provinces. This general equilibrium economic system allows us to demonstrate a new equilibrium with respect to an air pollutant reduction policy scheme. We focus on Chungcheongnam-do, where thermal power plants are concentrated in Korea, and examine an optimal policy to reduce particulate matter while minimizing the negative effects on the local economy. The rest of this study proceeds as follows. Section 2 elaborates the Korean air pollutant reduction policy, including the “Comprehensive Measures for Particulate Matter Management” and the current thermal power generation situation in Korea, where most production is focused in the Chungcheongnam-do province of Korea. Section 3 discusses the methods and data. Section 4 presents scenarios and discusses the effects of policy options. Concluding remarks and further discussions are in Section 5.

2. Current Policy and Case Study

2.1. Comprehensive Measures for Particulate Matter Management in Korea

Recently, the Korean government has strengthened its PM mitigation policies with specific policy measures mitigating not only PM itself, but also mitigating PM precursors such as NO$_x$, SO$_x$, volatile organic compounds (VOCs), and ammonia that contribute to secondary PM production [15]. The Korean government first set mitigation goals for PM precursors in the “Comprehensive Measures for Particulate Matter Management” in 2019. The mitigation goal has since been tightened to a national PM$_{2.5}$ goal of 16 $\mu$g/m$^3$ or below, with reductions to domestic emissions of 35% by 2024, through “The Special Law for Mitigating Particulate Matter (PM) Pollution” in 2019 [16]. That legislation defined the Korean PM pollution as a “social disaster” and greatly increased the national budget for urgent PM control policies, raising the budget by about 1.6 billion USD in 2019. In order to make this policy successful, the Korean government needs to prepare effective policy tools to control diverse domestic emission sources across the transportation, energy, and agriculture industries. In particular, Chungcheongnam-do (hereafter, CN) is known to have higher PM pollution due to its complex emission sources, including industry complexes (e.g., the Daesan industry complex), large coal power plants on its western coast (at Dangjin and Taeahn), and non-point emissions from agricultural areas (i.e., ammonia emissions from the livestock industry); see Figure 1 [17]. Thus, the Korean government has designated an additional three selected air pollution management areas that have higher emissions with higher annual PM concentrations in Korea. For instance, Chungcheongnam-do, one of the air pollution management areas, has to provide a regional plan for emission reductions, and its implementation actions need to be reviewed and evaluated regularly by the Ministry of Environment in Korea. These new regional targets are based on the recent “Basic Plan of Air Quality Management for Regional Air Pollution Management Area” [18]. This new regional strategy is very crucial for meeting the national PM reduction targets from the Comprehensive Measures for Particulate Matter Management, announced in 2019.

![Figure 1](image-url)  
**Figure 1.** Location of large point-source particulate matter (PM) pollutants (i.e., Daesan industry complex) and coal-power plants (at Dangjin and Taeahn) in Chungcheongnoam-do (CN).
2.2. Case Study: Chungcheongnam-Do and Korea

It is well known that fossil-fuel power plants are major sources of air pollutants, including particulate matter. For instance, coal-fired power plants and natural-gas-fired power plants emit SOx and NOx, including PM [19]. In Korea, there are four different types of power plants unevenly distributed across the nation. The Korean peninsula is geologically high in the east and low in the west, with most of the hydro-power plants located in the eastern parts of Korea (Gangwon-do), where ample natural fresh water flows most of the year, fed by mountain rains. Nuclear power plants need a massive water supply for cooling and large industrial complexes to insure the effectiveness of the generation and transmission of electricity. They are mostly located in the southern part of the peninsula (e.g., Gyeongsangbuk-do and Jeollanam-do). To meet peak-load electricity demand, thermal power plants are built in inland areas such as CN. The annual power generation in Korea is approximately 520,000 GWh, of which about 75% is supplied by thermal power plants. Of those thermal power plants, approximately 30% are located in CN, which has been experiencing severe air quality degradation. CN is responsible for about 22% of the national annual power generation, mostly from thermal power plants; see details in Figure 2 and Appendix A, Table A1 [20]. Moreover, industrial complexes (i.e., Daesan industry complex) and a manufacturing-based regional industrial geography may exacerbate the air conditions in CN. In terms of gross domestic product (GDP), the gross regional domestic product (GRDP) of CN is only about 6% [21] of the national total.

With a high concentration of thermal power plants in contrast to its low contribution to the GDP, CN’s annual emissions of PM10 and PM2.5 are almost 25% of the annual national emissions, accounting for approximately 31,000 tons of PM10 per year and 76,000 tons of PM2.5 per year of the Korea-wide totals [22]. Most of the PM10 and PM2.5 emissions come from major emission industries: agriculture, forestry, and fisheries; petroleum and coal; chemicals; non-metallic minerals; basic metals; electricity, gas, and steam supply; construction; and transportation. Of those industries, the basic metals industry is the largest PM polluter in CN, with more than 80% of the total PM emissions for CN. Non-metallic mineral products and transportation are the biggest PM-emitting industry sectors in Korea outside of CN, accounting for 96% of the PM10 and 97% of the PM2.5 total emissions in Korea, respectively. The PM emission contribution by industry in CN and the rest of the Korean provinces (hereafter, RP) can be found in Figure 3.
Two policy mechanisms for controlling PM emissions are allowed in the model—a PM$_{2.5}$ permits market and the pollution tax rates. The CGE model divides the economy into production, point-sources. Figure 4 clearly shows that CN suffers from almost year-round PM pollution.

According to Kim et al. [17], CN experiences severe air quality degradation, coupled with a high density of PM-polluting industries and large thermal power plants that are treated as large point-sources. Figure 4 clearly shows that CN suffers from almost year-round PM pollution.

3. Methodology and Data

3.1. Regional Computable General Equilibrium Model

Most air quality regulations have negative impacts on the economy in general. Respecting this, we developed a single-country regional static computable general equilibrium (CGE) model to analyze the economic effects associated with policies regulating air quality. Based on a traditional neoclassical CGE model [23], a single-country economy is divided into two parts, the CN and the remaining Korean provinces (RP), to compare the regional effects using a regional input–output table. Two policy mechanisms for controlling PM emissions are allowed in the model—a PM$_{2.5}$ emission permits market and the pollution tax rates. The CGE model divides the economy into production, household, government, savings and investment, and overseas sectors. The model consists of balanced equations that represent each sector’s optimal behavior and the flow of goods and production factors.
between sectors. In Figure 5, the production sector comprises a two-stage nesting structure. At the final level, domestic production is a composite of the value added and the intermediate input from Leontief technology. At the bottom level, capital and labor produce a composite of the value added in a constant elasticity of substitute (CES) technology.

\[
\begin{align*}
\text{Domestic Production (XD)} & \quad \text{Air pollutant coefficient (k)} \\
\text{Value Added (VA)} & \quad \sigma_F=0 \\
\text{Intermediates (I)} & \quad \sigma_{I}=0 \\
\text{Labor (L)} & \quad XD_1 \\
\text{Capital (K)} & \quad XD_2 \quad \ldots \quad XD_{i-1} \quad XD_i \\
\end{align*}
\]

**Figure 5.** Nesting structure of the production sector in a single-country regional static computable general equilibrium (CGE) model.

In the production sector, the demand functions for the input factor demand are derived through the producer’s cost minimization as Equations (1) and (2).

\[
\begin{align*}
\text{Min } C &= (1 + Tl_i) \cdot P_L \cdot L_i + (1 + Tk_i) \cdot P_K \cdot K_i, \\
\text{s.t. } XD_i &= aF_2(\gamma F_2 \cdot K_i^{-\sigma F_2} + (1 - \gamma F_2) \cdot L_i^{-\sigma F_2})^{-\frac{1}{\sigma F_2}},
\end{align*}
\]

where \( C \) is the cost function; \( XD_i \) is the final domestic production of \( i \) industry; \( L_i \) is the labor input; \( K_i \) is the capital input; \( P_L \) is the labor price; \( P_K \) is the capital price; \( TL_i \) is the labor tax rate; \( TK_i \) is the capital tax rate; \( aF_2 \) is the technical efficiency parameter; \( \gamma F_2 \) is the distribution parameter; and \( \sigma F_2 \) is the elasticity of substitution.

Various measures are possible for mitigating pollutants generated as by-products in the production process: direct regulation, an emissions tax, and a tradable permits market. The implementation of an emissions tax is practically difficult when the source of a pollutant is hard to identify. On the other hand, it is easier to measure outputs than emissions, so a tax on production can induce industries to reduce output and polluting emissions without additional information. When the tax imposed on the final product is increased, the zero-profit condition at the tax rate changes, but the production function remains unchanged in the CGE model. If, to reduce air pollutants, a tax is imposed in proportion to the final domestic production, the imposed tax \( t \) is added into the zero-profit condition at the final level of the nested production model, which is shown in Equation (3).

\[
PD_i \cdot XD_i \cdot (1 - tax) = (1 + Tk) \cdot P_K \cdot K_i + (1 + TL_i) \cdot P_L \cdot L_i + \left( \sum_{j=1}^{J} P_{j} \cdot iot_{j} \right) XD_i,
\]

where \( PD_i \) is the domestic price of \( i \) industry, \( P_j \) is the domestic price of other product of industry \( j \), and \( iot \) is the coefficient of intermediates inputs in the input-output table.

A system of marketable emissions permits is designed to induce the efficient use of resources by properly incentivizing polluters. An emissions trading system directly controls quantity, while an emissions tax controls price. Polluters can buy and sell their own permits to emit pollutants, and their emissions permits can be traded in the market. To ensure that the production cost increases by the number of permits purchased, air pollution permits are issued proportional to the amount of PM2.5 emissions that the firm releases in its (fixed-technology) production process. The free allocation of permits is modeled so that the firm receives subsidies from the government equal to the cost of...
purchasing the freely allocated permits. Therefore, as the tradable market is introduced, the company’s production function remains unchanged, but its cost function (Equation (4)) and zero-profit condition (Equation (5)) change.

\[
Cost = (1 + Ti)P_L L_i + (1 + Tk_i)P_K K_i + PPM_{2.5} ePM_{2.5},
\]

where \(PPM_{2.5}\) is the price of \(PM_{2.5}\) and \(ePM_{2.5}\) is the quantity of the \(PM_{2.5}\) emissions.

\[
PD_i XD_i(1 - tax^e) = (1 + Tk_i)P_K K_i + (1 + Ti)P_L L_i + \left( \sum_{j=1}^{J} P_j i o_{ij} \right) XD_i + PPM_{2.5} ePM_{2.5}
\]

The household sector represents all consumers, as suppliers of labor and capital for consumption, savings, and taxes. In addition, household income consists of labor and capital supply and government transfers. The demand functions for consumer goods are derived through the process of utility maximization under budget constraints, and a Cobb–Douglas function represents the household utility function in the model.

The model assumes that governments exist in the CN and the RP regions. The additional tax revenue from the air pollution reduction policy accrues to each local government and is assumed to be applied efficiently to offset damages—that is, for ecological conservation and the health of local residents. Government revenue consists of taxes on consumption, taxes on production, income taxes, and tariffs between the two regions. Similarly, government expenditures are the government consumption of consumer goods, labor and capital, and household transfers. The government converts the remaining amount (revenues minus expenditures) into savings. The demand functions for government consumption are derived through the process of utility maximization, again using a Cobb–Douglas utility function. Additional tax revenue is levied on thermal power generation in each region.

The total savings consist of household savings, government savings, and foreign savings. Total savings are invested by each industry, with each industry’s investment demand function derived through a utility maximization problem within its total savings. This study assumes that a Cobb–Douglas function is the utility function for the investment sector. The foreign sector is consumed domestically or exported with constant elasticity of transformation (CET) technology under the Armington assumption. The model structure can be found in Figure 6.

Figure 6. Structure and interwoven relationships among the agencies of the regional computable general equilibrium model (note: blue squares are two policy schemes used in the model and arrows indicate the flow of economic activities).
3.2. Sector Classification

Regional CGE models are mainly used to analyze regional economies, economic development, and policy options. CGE models are built either by applying Armington’s [24] assumptions about transactions among regions, or by incorporating a complete CGE model for each region as if the total national economy were a global model [25]. In this research, a regional CGE model was constructed using an industrial classification to establish a framework for a regional input–output table and to exclude unnecessary assumptions. A social accounting matrix (SAM) defines transactions among industries and other agents and is used to calibrate the CGE model. The calibrating SAM for this project is based on a Bank of Korea regional input–output table [26], with 2013 taken as the base year. We classify 60 industries (i.e., 30 industries, 2 regions), with industries in CN between 1 and 30, and those in RP between 31 and 60. The elasticity of substitution between labor and capital at the stage of value-added production is applied the same, regardless of region. Details of the industry classification and PM emissions can be found in Table A2.

In this research, the PM$_{2.5}$ emissions in the base year are classified by the same SAM criteria in order to analyze changes in the PM$_{2.5}$ emissions. While carbon emissions are released in proportion to fossil-fuel combustion, PM$_{2.5}$ emissions are less likely to be highly correlated with CO$_2$ emissions, so air pollutant emissions are defined in proportion to the industrial final production. Air pollutant emissions per value of domestic production ($g$/USD) are obtained by dividing each quantity of emissions in grams by the regional domestic production in USD. Thus, each industry has its own emissions coefficient for PM$_{2.5}$. The coefficients represent the emission intensity, and are calculated by pairing the national air pollutant emission figures [22] with the regional input–output table industries [26,27]. Figure 7 shows that the emissions from basic metal products and transportation are particularly high in CN, while non-metallic products and transportation are high in RP (see Appendix A, Table A2, for more details).

![Figure 7. PM$_{2.5}$ emission per domestic production for the 8 highest emission industries in CN and RP.](image)

4. Scenario Building and Results

4.1. Scenarios

It is well known that an emissions tax is more efficient and effective than most means of direct regulation in mitigating air pollution. However, a direct regulation of pollutants may be more cost-effective to implement. In fact, the main policy schemes in Korea for mitigating air pollution of the PM type are direct regulations. For instance, the Korean government announced regulation policy measures to reach national air standards by shutting down aging coal-fired power plants and enforcing upper limits to power generation when the concentrations of PM$_{10}$ and PM$_{2.5}$ are particularly high [15]. To analyze the policy effects on the Korean economy, the aforementioned policy schemes can be treated as spontaneous events (external shocks) to an economic structure. In this
way, our experiment is focused on the long-term economic effects and environmental impacts of air quality regulation—e.g., in response to a regional development tax in Korea. The purpose of this tax, imposed as a proportion of power generation, is to secure funds for balanced regional development, improvements to environmental quality, and the protection of water resources.

The current tax is 0.03 cents per kWh for thermal power, 0.2 cents per kWh for hydro power, and 0.1 cents per kWh for nuclear power. However, there has been increasing pressure to raise the tax rates for thermal power, equivalent at least to the level of water or nuclear power. The revenue from this tax could be used to improve the pollution mitigation capacities of thermal power plants, lowering emissions that affect local ecosystems and local residents [28]. If this regional development tax rate increases to the nuclear power rate, the long-term economic and environmental impacts could be more permanent. With this in mind, we simulate the permanent impacts on regional economies and the environment by executing a sensitivity analysis, ranging the tax from the status quo (0.03 cents per kWh) up to 1 cent per kWh. In the scenarios, the regional development tax impacts the electricity, gas, and steam supply industry in Table 1, which is industry 16 for CN and industry 46 for RP, and the ratio of the tax is recalculated for each scenario.

| Industry Classification | $\sigma F_2^1$ | Industry Classification | $\sigma F_2^1$ |
|------------------------|----------------|------------------------|----------------|
| 1; 31 Agricultural, forestry, and fishery goods | 1.014 | 16; 46 Electricity, gas, and steam supply | 1.004 |
| 2; 32 Mined and quarried goods | 0.902 | 17; 47 Water supply, sewage, and waste management | 0.992 |
| 3; 33 Food, beverages, and tobacco products | 0.995 | 18; 48 Construction | 0.992 |
| 4; 34 Textile and leather products | 1.005 | 19; 49 Wholesale and retail trade | 0.999 |
| 5; 35 Wood and paper products, printing, and reproduction of recorded media | 0.998 | 20; 50 Transportation | 1.009 |
| 6; 36 Petroleum and coal products | 1.101 | 21; 51 Food services and accommodation | 0.971 |
| 7; 37 Chemical products | 0.999 | 22; 52 Communications and broadcasting | 1.01 |
| 8; 38 Non-metallic mineral products | 0.998 | 23; 53 Finance and insurance | 1.004 |
| 9; 39 Basic metal products | 0.976 | 24; 54 Real estate and leasing | 0.931 |
| 10; 34 Fabricated metal products, except machinery, and furniture | 1.004 | 25; 55 Professional, scientific, and technical services | 0.995 |
| 11; 34 Machinery and equipment | 0.985 | 26; 56 Business support services | 1.011 |
| 12; 34 Electronic and electrical equipment | 1.001 | 27; 57 Public administration and defense | 1.001 |
| 13; 34 Precision instruments | 1.001 | 28; 58 Educational services | 1.002 |
| 14; 34 Transportation equipment | 0.985 | 29; 59 Health and social work | 0.987 |
| 15; 45 Other manufactured products and outsourcing services | 1.002 | 30; 60 Cultural and other services | 0.999 |

In addition, an emission permits market is designed to provide economic incentives to polluters, similar to an emissions tax, an emissions charge, or an emissions trading system to control the quantity of pollutants [29]. In this research, we introduce a hypothetical PM emission permits market system to analyze whether this policy scheme can reduce the particulate matter in Korea or not, and to analyze what the total economic impact will be, whether negative or positive. In total, we provide seven different combinations of scenarios to calculate the regional economic impacts, and from this set determine the optimal policy for reducing PM in Korea. Scenarios used in this research can be found in Table 2.
Table 2. Scenarios for combined air quality improvement policies.

| Scenario          | Description                                                      | Tax Rate (Cent/kWh) | Permit Market |
|-------------------|------------------------------------------------------------------|---------------------|---------------|
| Status quo (baseline) | Current situation                                                | 0.03                | No            |
| Current + M       | Implementing permit market only                                   | 0.03                | Yes           |
| Hyd               | Applying regional development tax of hydropower                   | 0.20                | No            |
| Hyd + M           | Applying regional development tax of hydropower and implementing permit market simultaneously | 0.20                | Yes           |
| Mod               | Applying regional development tax higher than hydropower          | 0.50                | No            |
| Mod + M           | Applying regional development tax higher than hydropower and implementing permit market simultaneously | 0.50                | Yes           |
| Stro              | Applying regional development tax significantly higher than hydropower | 1.00                | No            |
| Stro + M          | Applying regional development tax significantly higher than hydropower and implementing permit market simultaneously | 1.00                | Yes           |

4.2. Economic Impact

Under the current regional development tax scheme, the GDP in Korea is approximately 1303 billion USD. Of this number, CN accounts for 81 billion USD, and RP accounts for 1222 billion USD. Although the national GDP increases gradually as the regional development tax rises, the GRDP in CN slightly decreases by 0.04% under the alternative “tax rate as for water power plant” scenario—i.e., 0.2 cents/kWh without introducing a permits market system. In this scenario, the government expenditure increases by approximately 2.05%, but household consumption and investment decrease 0.17% and 0.07%, respectively. It is assumed that government expenditure is not enough to compensate the consumption decreases. Moving from the “moderate” to the “strong” regulation scenario—one imposing a 0.5 cent/kWh and 1 cent/kWh tax still without a permits market—the GRDP increases for CN and decreases for RP. However, the increase in the CN GRDP more than offsets the decrease in the RP GRDP, leading us approximately 1303 billion USD in the national GDP, a 0.24% increment compared to the status quo.

When a PM$_{2.5}$ emission permits market is introduced in the model, the government revenue increases, and the new revenue is assumed to be reinvested in the clean energy sector locally. In the last of the seven scenarios, “stro + M (strong regulation with permits market),” we discover the optimal result, maximizing the aggregate economic impact nationally at 1303.26 USD. The 81.39 billion USD for the CN is the highest across the scenarios at +0.48% over the status quo, more than enough to offset a −0.028% loss in the GRDP for RP. Table 3 presents a detailed comparison of the GRDP changes under each scenario for CN and for RP.

Table 3. Comparison of GRDP under different scenarios for CN and RP.

| Scenario                      | Status Quo | Hyd | Mod | Stro | Hyd + M | Mod + M | Stro + M |
|-------------------------------|------------|-----|-----|------|---------|---------|---------|
| National GDP (billion USD)    | 1303.20    | 1303.21 | 1303.22 | 1303.24 | 1303.22 | 1303.24 | 1303.26 |
| GRDP in CN (billion USD)      | 81.00      | 80.96 | 81.02 | 81.20 | 81.12   | 81.21   | 81.39   |
| % change in GRDP              | −0.04      | +0.03 | +0.24 | +0.15 | +0.27   | +0.48   |         |
| % change in household         | −0.17      | −0.48 | −1.01 | −0.21 | −0.56   | −1.18   |         |
| GRDP in RP (billion USD)      | 1222.21    | 1222.24 | 1222.20 | 1222.04 | 1222.10 | 1222.02 | 1221.87 |
| % change in GRDP              | +0.0032    | −0.0004 | −0.0135 | −0.0088 | −0.0150 | −0.0277 |         |
| % change in household         | −0.08      | −0.22 | −0.48 | −0.10 | −0.27   | −0.58   |         |
| % change in government        | +0.31      | +0.86 | +1.80 | +0.44 | +1.18   | +2.51   |         |
| % change in investment        | −0.07      | −0.20 | −0.43 | −0.13 | −0.35   | −0.76   |         |

4.3. Environment Impact

The PM$_{2.5}$ emissions in the CGE model are calculated by multiplying the emission intensity by the total production. As the tax rate increases from 0.2 cents to 1 cent, the national total emissions decrease from 0.1% to 0.64%, compared to emissions at the current tax rate. There is a difference between CN and RP in the reduction pattern with respect to industry classification. The top five PM$_{2.5}$-emitting industries (basic metal products; electricity, gas, and steam supply; construction; transportation; and non-metallic mineral products) were defined as high-emission industries under...
the status quo. In the current situation, the top five industries account for 98% of the PM$_{2.5}$ emissions in CN and 91% in RP. The decrease in CN is larger than that in RP, and these regional differences were especially noticeable in the five high-emission industries. With the tax rate rising from 0.2 cents to 1 cent, the emissions in CN decrease between 0.19% and 1.11%, while they decrease only between 0.08% and 0.50% in RP. The reductions for the five high-emission industries are greater than the reduction for the other 25 industries in both CN and RP in every scenario.

The introduction of a PM permits market has a positive effect on the environmental performance when combined with an increase in the tax rate. As the tax rate increases from 0.2 cents to 1 cent, the national total emissions decrease from 0.44% to 2.35% with a tradable market, specifically from 0.65% to 3.53% in CN and from 0.37% to 1.99% in RP. The presence of an emissions market varies depending on the industry clustering in both CN and RP. In the case of the five high-emissions industries, the reduction is about three times in CN and about four times in RP compared to the status quo scenario. The reduction for the remaining 25 industries due to the presence of the market is relatively small. Compared to Oh et al. [6], the costs of reducing the PM air pollution are relatively small in our simulation. A comparison between CN and RP in terms of PM$_{2.5}$ emissions under different scenarios is shown in Table 4.

| Scenario                        | Status Quo | Hyd   | Mod    | Hyd + M | Mod + M | Stro + M |
|---------------------------------|------------|-------|--------|---------|---------|----------|
| National emissions (tons)        | 75,565     | 75,486| 75,339 | 75,081  | 75,234  | 74,741   | 73,791   |
| Total % changes                  |            | −0.10 | −0.30  | −0.64   | −0.44   | −1.09    | −2.35    |
| Emissions in CN (tons)           | 17,626     | 17,593| 17,534 | 17,431  | 17,511  | 17,336   | 17,004   |
| Total % change                   |            | −0.19 | −0.52  | −1.11   | −0.65   | −1.64    | −3.53    |
| % change in 5 high-emission      | 17,383     | −0.19 | −0.53  | −1.12   | −0.66   | −1.67    | −3.58    |
| industries                       |            |       |        |         |         |          |          |
| % change in the 25 other industries | 243       | −0.05 | −0.12  | −0.24   | −0.01   | −0.07    | −0.16    |
| Emissions in RP (tons)           | 50,939     | 57,893| 57,805 | 57,650  | 57,724  | 57,405   | 56,787   |
| Total % change                   | −0.08      | −0.23 | −0.50  | −0.37   | −0.92   | −1.99    | −2.14    |
| % change in 5 high-emission      | 52,203     | −0.08 | −0.24  | −0.51   | −0.40   | −0.99    | −2.14    |
| industries                       | 5236       | −0.06 | −0.17  | −0.38   | −0.08   | −0.21    | −0.43    |
within the market and regulatory (pricing and tax) system. CN in Korea was selected as a representative case study due to its high dependency on thermal power plants and industrial complexes. Instead of a severe PM mitigation action plan such as shutting down power plants, two economically flexible policy experiments were simulated: a regional development tax imposed as a proportion of power generation, and a tradable PM$_{2.5}$ emission permits market similar to a carbon trading market. Four different regional development tax rates are considered as scenarios: 0.03, 0.2, 0.5 and 1 cent per kilowatt hour. These tax rates are applied across the entire nation, and the tax revenues in CN are greater than those in RP, due to CN’s high concentration of thermal power plants. The model assumes that the new government tax revenue is reinvested in clean energy sectors. Owing in part to this effect, the GRDP in CN increases when the Korean government imposes higher tax rates on fossil-fuel energy production. Our simulation results offer a superior solution when a tradable PM$_{2.5}$ emissions market is introduced simultaneously. From an environmental perspective, the aggregate PM$_{2.5}$ emission reduction in the country is approximately 74,000 tons annually, and high PM emission industries in CN have to reduce their PM emission dramatically compared to those in RP under the optimal solution. This simulated number could be converted into an auxiliary benefit—a reduction in health risk.

We believe that by navigating hypothetical combined solutions, this research provides useful alternative policy schemes to reform ambient air quality, but concede that several modifications to our model could improve future investigations. First, introducing the other main sources of PM emissions, such as industrial complexes and the agricultural sector, would better reflect reality. Further, diverse emission source dynamics could introduce a new dimension beyond what was done in this research. Second, the secondary PM components and ammonia, other important byproducts of fossil-fuel energy sources, are ignored to simplify this model. Considering the biogeochemical reaction of ammonia in the atmosphere may lead us to a different policy path for mitigating air pollutants. Third, this study does not calculate the indirect costs of auxiliary benefits from reducing health risks linked to PM emissions, hence a formal cost-benefit analysis of implementing air quality improvement plans could lead to a new optimum economic equilibrium. Finally, since particulate matter can be treated as a transboundary pollutant, a broader research area including at least northeast Asia should be investigated to better simulate real-world conditions. We believe that our first attempt to study mitigation policies for air pollutants by introducing inter-regional economic differences is a good starting point and provides some guidance for how to steer promising research in the future.

**Author Contributions:** K.P. and Y.L. contributed to the design and development of this research; K.P. developed the program coding for a static regional CGE model; C.S., T.Y. and Y.H. provided policy guidelines to build policy scenarios for the model; K.P. and Y.L. wrote the bulk of the manuscript; E.K. provided GIS graphics and analysis. All authors have read and agreed to the published version of the manuscript.

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**Conflicts of Interest:** The authors declare no conflict of interest.
### Appendix A

#### Table A1. Power generation by regions.

| Region           | Hydro | Thermal | Nuclear | Renewable | Total |
|------------------|-------|---------|---------|-----------|-------|
| Seoul            | 0     | 29      | 0       | 41        | 70    |
| Busan            | 0     | 842     | 1287    | 41        | 2170  |
| Daegu            | 0     | 250     | 0       | 9         | 259   |
| Incheon          | 4     | 5803    | 0       | 42        | 5849  |
| Gwangju          | 1     | 60      | 0       | 8         | 69    |
| Daejeon          | 0     | 38      | 0       | 1         | 39    |
| Ulsan            | 0     | 1747    | 1106    | 4         | 2857  |
| Gyeonggi-do      | 56    | 7333    | 0       | 124       | 7513  |
| Gangwon-do       | 115   | 1832    | 0       | 220       | 2167  |
| Chungcheongbuk-do| 39    | 35      | 0       | 38        | 112   |
| Chungcheongnam-do| 6     | 11,650  | 0       | 631       | 12,287|
| Jeollabuk-do     | 53    | 862     | 0       | 109       | 1024  |
| Jeollanam-do     | 4     | 2405    | 3102    | 416       | 5927  |
| Gyeongsangbuk-do | 130   | 200     | 4943    | 334       | 5607  |
| Gyeongsangnam-do | 58    | 5139    | 0       | 47        | 5244  |
| Jeju-do          | 0     | 232     | 0       | 81        | 313   |
| Sejong           | 0     | 342     | 0       | 3         | 345   |
| Total            | 466   | 38,799  | 10,438  | 2149      | 51,852|

Sources: Korea Electric Power Corporation [18].

#### Table A2. PM emission and emission intensity by industries in 2013 (g/1000 USD).

| Industry Classification | Emissions (Tons/Year) | Emission Intensity |
|-------------------------|-----------------------|--------------------|
|                         | PM₁₀ | PM₂.₅ | PM₁₀ | PM₂.₅ |
| 1; 31 Agricultural, forest, and fishery goods | 207.5 | 180.1 | 23.64 | 1575.8 | 1339.0 | 27.43 |
| 2; 32 Mined and quarried goods | 0.2 | 0.1 | 0.37 | 81.5 | 29.2 | 7.31 |
| 3; 33 Food, beverages and tobacco products | 1.2 | 1.0 | 0.09 | 735.0 | 263.2 | 2.87 |
| 4, 34 Textile and leather products | 0.3 | 0.2 | 0.12 | 279.0 | 99.2 | 1.31 |
| 5, 35 Wood and paper products, printing and reproduction of recorded media | 0.8 | 0.6 | 0.21 | 189.0 | 78.0 | 2.21 |
| 6, 36 Petroleum and coal products | 47.0 | 13.4 | 0.59 | 2487.7 | 1266.2 | 9.51 |
| 7, 37 Chemical products | 39.3 | 26.9 | 0.69 | 1699.8 | 1123.6 | 4.95 |
| 8, 38 Non-metallic mineral products | 34.0 | 28.1 | 4.50 | 43,077.1 | 22,508.1742.94 |
| 9, 39 Basic metal products | 130 | 200 | 14,213.3 | 479.25 | 9958.3 | 6056.5 | 33.39 |
| 10, 40 Fabricated metal products, except machinery and furniture | 1.0 | 0.8 | 0.11 | 398.4 | 148.2 | 1.71 |
| 11, 41 Machinery and equipment | 0.6 | 0.5 | 0.06 | 281.0 | 102.6 | 0.93 |
| 12, 42 Electronic and electrical equipment | 4.3 | 3.6 | 0.06 | 569.4 | 231.2 | 0.75 |
| 13, 43 Precision instruments | 0.3 | 0.2 | 0.11 | 178.7 | 64.3 | 2.67 |
| 14, 44 Transportation equipment | 1.4 | 1.2 | 0.06 | 418.9 | 157.8 | 0.71 |
| 15, 45 Other manufactured products and outsourcing services | 1.8 | 1.3 | 0.33 | 150.0 | 58.7 | 1.07 |
| 16, 46 Electricity, gas, and steam supply | 1883.1 | 1376.3 | 130.36 | 2515.6 | 2135.5 | 23.83 |
| 17, 47 Water supply, sewage and waste management | 12.8 | 10.8 | 6.12 | 229.8 | 191.2 | 9.29 |
| 18, 48 Construction | 440.0 | 404.8 | 22.70 | 5755.7 | 5295.3 | 30.65 |
| 19, 49 Wholesale and retail trade | 0.7 | 0.5 | 0.10 | 26.5 | 17.1 | 0.07 |
| 20, 50 Transportation | 1483.0 | 1364.3 | 377.72 | 18,149.4 | 16,697.4 | 129.40 |
| 21, 51 Food services and accommodation | 0.6 | 0.4 | 0.09 | 19.9 | 12.8 | 0.14 |
| 22, 52 Communications and broadcasting | 0.3 | 0.2 | 0.13 | 9.8 | 6.3 | 0.06 |
| 23, 53 Finance and insurance | 0.5 | 0.3 | 0.12 | 18.9 | 12.2 | 0.09 |
| 24, 54 Real estate and leasing | 0.3 | 0.2 | 0.07 | 10.8 | 7.0 | 0.05 |
| 25, 55 Professional, scientific, and technical services | 0.3 | 0.2 | 0.06 | 11.7 | 7.5 | 0.07 |
| 26, 56 Business support services | 0.5 | 0.3 | 0.25 | 17.3 | 11.1 | 0.27 |
| 27, 57 Public administration and defense | 0.0 | 0.0 | 0.006 | 1.6 | 1.0 | 0.01 |
| 28, 58 Educational services | 0.0 | 0.0 | 0.002 | 0.6 | 0.4 | 0.004 |
| 29, 59 Health and social work | 0.1 | 0.0 | 0.013 | 2.6 | 1.7 | 0.017 |
| 30, 60 Cultural and other services | 0.3 | 0.2 | 0.05 | 11.6 | 7.5 | 0.10 |

Sources: National Center for Fine Dust Information [20].
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