Disinfection and reduction of organic load of sewage water by electron beam radiation

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Abstract The efficacy of electron beam radiation for the disinfection and reduction of organic load of sewage water was assessed with ILU-6 Accelerator at Radiation Technology Development Division of the Bhabha Atomic Research Centre, Mumbai India. The current problem on environmental health in relation to water pollution insists for the safe disposal of sewage water. In general, sewage water comprises heterogeneous organic based chemicals as well as pathogens. EB treatment of the wastewater has found to be very effective in reducing the pathogens as well as organic load. EB dose of 1.5 kGy was sufficient for complete elimination of total coli forms. The experimental results elucidated the reduction of biological oxygen demand—BOD (35 and 51.7%) in both inlet and outlet sewage samples. Similarly reduction of chemical oxygen demand—COD was observed (37.54 and 52.32%) in both sewage samples with respect to increase in irradiation doses (0.45–6 kGy). The present study demonstrated the potential of ionizing radiation for disinfection of sewage and to increase the water quality of the wastewater by decreasing BOD and COD. So, the irradiation sewage water can find its application either in agriculture for irrigation, in industry for cooling purpose and some selected domestic purposes.

Keywords Coliform and organic load of sewage water · Electron beam radiation

Introduction

Population and development pressures have made waste water treatment an important component of environmental pollution control alleviation. In India, major cities generate large volumes of wastewater due to their inefficient water management systems. Large amounts of wastewater that can be reused, is let down into either sewerage drains or nearby aquatic resources. Furthermore, the potential for environmental pollution is greater due to high quantity of waste materials produced namely the organic matter (Shathele 2009), generally rich in pathogens of faecal origin, even after the treatment with conventional methods. Few important pathogens in sewage water namely bacteria (E. coli, Salmonellae sp.), viruses (e.g. Poliovirus) and protozoans are causing various types of diseases especially of faecal oral route (Jianlong Wang and Jiazhuo Wang 2007). Therefore, to protect health and the environment, new approaches should be adopted to overcome this grim situation, including remediation and the reuse of wastewater. Although it is under scrutiny, chlorine is still widely used as chemical disinfectant for the treatment of wastewater. Treated water discharged from sewage treatment plant can be reused for irrigation. Prescribed water quality standards for recreational and other water uses, such as shellfish production, are typically based on contamination levels as indicated by number of faecal coliforms in effluents. Recently, application of ionizing irradiation, especially electron beam radiation that is generated from electron beam accelerator to treat sewage water has attracted the environmentalists. Chief advantage of
electron beam radiation over existing chemical treatments derives from its short wavelength and its ability to penetrate into the tissues. It was reported that ionizing radiation induces both the degradation of numerous compounds and the inactivation of microorganisms in sewage on irradiation, depending on the type of energy, dose rate and absorbed radiation (Klieber et al. 2002; Drzewicz and Gehringer 2005; Casimiro et al. 2008).

In water as well as wastewater, the principal component is water. Therefore, it would be expected that the effect of ionizing radiation might be dominated by interaction of radiation and water (Meeroff Daniel et al. 2004). As far as pure water is concerned, when exposed to ionizing radiation, the radiolysis of water can be presented as following equation (Spinks and Woods 1990):

\[
\begin{align*}
H_2O & \rightarrow [2.7]OH^+ + [2.6]e_{aq}^- + [0.6]H^+ + [2.6]H_2O^+ \\
& + [0.7]H_2O_2 + [0.45]H_2
\end{align*}
\]

So if we consider pure water, each 100 eV absorbed by water will result in the generation of 2.7 radical OH*, 2.6 e_{aq}^-, 0.6 radical H*, 2.6H_2O^+, 0.7 molecule of H_2O_2 and 0.45 molecule of H_2 (Borrely et al. 1998a, b).

Powerful oxidizing and reducing species (e.g., OH, e_{aq}^-, H) and molecular products (e.g., H_2, H_2O_2) are produced due to the interaction between radiation and water, so these chain reactions lead to the phenomenon as described earlier by Parker and Darby (1995) and Sommers and Glenn (2006).

Initial research at the electron beam facility focused on bacterial disinfection at the large scale (Farooq et al. 1993) as well as the effect of electron beam irradiation on waste water characteristic such as biological oxygen demand (BOD) and chemical oxygen demand (COD). Therefore, the effects of electron beam irradiation on the inactivation of microorganisms, mainly on the pathogens, as well as reduction of organic load present in sewage water were investigated.

Methods

The water samples were collected from sewage water treatment plants (A & B), Mumbai (inlet area from where raw sewage is entering into the plant and outlet i.e. prior to chlorination). The samples were brought to the laboratory with the help of sterile plastic bottles, kept in ice boxes and analysed for BODs, COD and total microbial count by adopting standard procedures (APHA 2005).

Irradiation of sewage water

A laboratory scale study has been carried out regarding the possibility of electron beam application for sewage water disinfection and reduction of BOD and COD. To carry out the experiment, the stainless steel belt conveyor was constructed for irradiation under continues flowing condition. The width and depth of water container was designed to be 60 and 0.7 cm, respectively, which was fit to the penetration range of experimental accelerator. The rate of sewage water moving through conveyor belt was controlled in ~3,000 cc/min, which treat up to 1.5 m³/h. The radiation experiment was carried using ILU-6 Accelerator (ILU-6 pulse Accelerator parameters are shown in Table 1), at Electron Beam Facility of the Radiation Technology Development Division of the Bhabha Atomic Research Centre, Mumbai, India. Then all samples were analysed for the same parameters after EB treatment at different doses (0.45, 0.75, 1.5, 3, 4.5 and 6 kGy) at room temperature.

Microbial count

Plate count method was used to enumerate the microbial count. The Hi-Media Plate count Agar (M091-500G) was used for total bacterial count, Hi-Media MacConkey Agar (M081A) for coliforms and Hi-Media Salmonella–Shigella Agar (M108) for total Salmonella–Shigella. The wastewater samples were diluted in Phosphate buffer (Hi-Media M 461) and were used on spread plate method on agar, then incubated at 37°C until colony formation was observed.

Results and discussions

The effect of electron beam radiation on coliform and the decomposition of organic materials in sewage water were evaluated as a function of absorbed dose by determining the change of microbial load, biological oxygen demand and chemical oxygen demand. As shown in Figs. 1, 2, 3, 4, 5, 6, 7 and Table 2.

EB treatment act as disinfectant

The mechanism of killing the pathogens depends largely on the nature of the disinfectant and on the type of microorganism. In general, four mechanisms are proposed to

| Table 1 | ILU-6 pulse accelerator parameters during the experiment |
|---------|--------------------------------------------------------|
| Energy  | 2 MeV                                                  |
| Average beam current | 1.2 mA                                               |
| Pulse repetition frequency | 10 Hz                                                 |
| Conveyor speed | ~ 10 cm/s                                               |
| Pulse current | 250 mA                                               |
| Pulse accelerator duration | ~ 500 μS                                           |

The dosimetry study was carried out using FWT60 radio chromic films calibrated by Graphite Calorimeter.
explain the destruction or inactivation of organisms, damage to cell well, alteration of cell permeability, changing the colloidal nature of the cell protoplasm and inactivation of critical enzyme systems responsible for metabolic activities.

Several samples were obtained from the plant to examine the inactivation of microorganisms with radiation technology. Figures 1, 2, 3, and 4 show the inactivation of total and faecal *coli*forms (organisms per ml) in the raw inlet and unchlorinated outlet water samples. Our study observed that both total and faecal *coli*forms were reduced to below detection limit. The total bacterial count reduced from $6.1 \times 10^8$ to 0 at 3.0 kGy, total *coli*forms count reduced from $9.6 \times 10^4$ to 3 at 0.75 kGy and total *Salmonella–Shigella* count reduced from $4 \times 10^3$ to 0 at 0.75 kGy in the case of treatment plant-A, inlet sewage water sample and total bacterial count reduced from $4.8 \times 10^4$ to 0 at 3.0 kGy, total *coli*forms count reduced from $1.6 \times 10^4$ to 1 at 1.5 kGy and total *Salmonella–Shigella* count reduced from $5 \times 10^3$ to 0 at 0.75 kGy.
Similarly for inlet sample of treatment plant-B, the total bacterial count reduced from $6.8 \times 10^8$ to 0 at 3.0 kGy, total *coliforms* count reduced from $10.6 \times 10^4$ to 36 at 0.75 kGy and total *Salmonella–Shigella* count reduced from $1.4 \times 10^4$ to 23 at 0.75 kGy and for outlet samples total bacterial count reduced from $5.2 \times 10^4$ to 0 at 3.0 kGy, total *coliforms* count reduced from $1.9 \times 10^4$ to 1 at 1.5 kGy and total *Salmonella–Shigella* count reduced from $5.4 \times 10^3$ to 13 at 0.75 kGy (Figs. 3, 4).

Radiation effects on microorganisms are not only associated with the chemical changes but are also dependent on physical parameters such as dose rate, dose distribution,
radiation quality, radiation type and exposure pattern and physiological factors like growth phase, sensitivity, number of microorganisms etc. The action of radiation on a living organism can be divided into direct and indirect effect. If the radiation interacts with the atoms of the DNA molecule or some other cellular component critical to the survival of the cell, it is a direct effect, which will eventually affect the ability of the cell to reproduce and survive. The formation of major radiolysis products from water and their subsequent interactions with waste particles are described as indirect effect, which is generally caused by energy deposition in the medium resulting in the formation of secondary reactants generated through free radicals production, sensitizer and secondary ionization (Borrely et al. 1998a, b).

The gross appreciation of radiation effects on microorganisms is due to two main parameters. (a) The lethal dose at which all the microorganisms are killed or inactivated. (b) The $D_{10}$ value that corresponds to the dose at which the number of viable microorganisms present in the initial medium is decreased by 1log. The radiation dose required to inactivate the pathogenic bacteria is generally defined in terms of the $D_{10}$ value, which is the radiation dose required to reduce (through inactivation or cell death) the microbial concentration by a factor of ten or by one log cycle. Infect, which is the basis principle for producing radiation-sterilized, single-use medical products, is now well established in industries worldwide (Pikaev and Woods 1994). Based on the same principle, the pathogens present in sewage water can also be effectively removed by exposure to high energy radiation. In this study, EB treatment was found to be very effective in reducing the bacterial count, especially total coliforms count to a safe level and the $D_{10}$ values of the Total Bacterial Count, Total Coliforms and Total Salmonella–Shigella was found to be $\leq 0.4$ kGy (Table 2) and 1.5 kGy was sufficient for 5 log elimination of total coliforms. (As shown in Figs. 1, 2, 3, 4) The effectiveness of gamma and EB

| S.No | Organisms                  | $D_{10}$ value |
|------|---------------------------|----------------|
| 1    | Total bacterial count     | $\leq 0.4$     |
| 2    | Total coliforms           | $\leq 0.4$     |
| 3    | Total Salmonella–Shigella | $\leq 0.4$     |
in eliminating bacterial load was also reported by Basfar and Abdel Rehim (2002) and Rawat et al. (1998).

**Sewage water quality improvement by EB treatment**

**Biological oxygen demand (BOD)**

The samples were irradiated with different doses ranging from 0.45 to 6.0 kGy. The removal percentage of BOD\(_5\) of inlet sewage water sample was increased from 11.11 to 35 for the dose 0.45–6.0 kGy and for outlet water samples the removal of BOD was up to 51.7% in the case of treatment plant-A as shown in Figs. 5 and 6. Similarly for treatment plant-B, the percentage of BOD\(_5\) was found to be increased from 7.73 to 30.38 in the case of inlet sewage water sample and for outlet water sample it was up to 47.16% with the same doses as represented in Figs. 7 and 8.

**Chemical oxygen demand (COD)**

In the present study, it was observed that the rate of COD removal of inlet sewage water sample was increased from 13.20 to 37.54 for the dose 0.45 to 6.0 kGy and for outlet water samples the removal of COD was up to 34.54% in the case of treatment plant-A as shown in Figs. 5 and 6. Similarly for treatment plant-B, the removal percentage of COD increased from 12.11 to 36.05 for the dose 0.45 to 6.0 kGy in the case of inlet sewage water samples and for outlet water samples the removal of BOD is up to 52.32% with the respective doses (Figs. 7, 8).

Biological oxygen demand and chemical oxygen demand are the major parameters used as routing surrogate tests for measuring the load of organic carbon into the water environment. In pollution abatement works, organic matter measurement provides a quick and simple method for monitoring pollution levels as well as the accumulation of non-biodegradable or refractory organic materials. The

**Table 3** Preliminary and rough cost estimation and comparison for disinfection of sewage water (US $/m³)

| M³/day | Electron beam | Ozone | UV | Chlorine |
|--------|---------------|-------|----|----------|
| 10,000 | 0.29          | 0.25  | 0.17| 0.013    |
| 50,000 | 0.073         | 0.086 | 0.171| 0.013    |
| 100,000| 0.050         | 0.064 | 0.054| 0.013    |
| 200,000| 0.041         | 0.053 | 0.047| 0.013    |

US $/m³ Total cost including operation, capacity and discharge water in eliminating bacterial load was also reported by Basfar and Abdel Rehim (2002) and Rawat et al. (1998).

**Table 4** Qualitative comparison of various treatment systems

| Factor considered       | Activated sludge plant | Extended aeration activated sludge | Biological filter | Oxidation ditch | Aerated lagoon | Chlorination for disinfection | Radiation with electron beam accelerator |
|-------------------------|------------------------|------------------------------------|-------------------|-----------------|----------------|-------------------------------|------------------------------------------|
| Organic matter removal  | F                      | F                                  | F                 | G               | G              | –                             | G                                        |
| FC removal              | P                      | F                                  | P                 | F               | G              | G                             | EX                                       |
| Economic factors        | L                      | L                                  | L                 | M               | M              | M                             | H                                        |

Adapted from Pereira et al. 2002

FC faecal coli form, EX excellent, G good, F fair, P poor, H high, M medium, L low (e.g. low demanding or low cost)
efficiency of waste water treatment plants and kinetics characteristics of biological process are usually studied through BOD and COD (Aziz and Tebbutt 1980; Fadini et al. 2004 and Ali Vahdat et al. 2010). In aqueous media, the oxidizing hydroxyl radical OH*, the reducing hydrated electron eaq- and the hydrogen radical H* are the predominant products, all of which are highly reactive transient species and are responsible for the various effects including the reduction of pathogens, the oxidation of hazardous organic pollutants and the destruction of molecular structure of targeted pollutants. (Maruthi et al. 2011). The effectiveness of ionizing radiation to remove toxic organic chemicals from aqueous solutions has been already reported by Huaying Bao et al. (2002) and Getoff (1999). The present study revealed that the reduction percentage of BOD and COD of outlet samples was more than the raw sewage samples with same dose (6 kGy) because of other interference like turbidity, total solids and heavy metals etc. which were present in raw sewage before conventional treatment. By the present study it was observed that some of the non-biodegradable fractions got converted into biodegradable fraction as indicated by increase in BOD/COD with an increased treatment dose and it was incredibly effective when sewage contain more organic and inorganic load (shown as Fig. 9).

Moreover, treatment of sewage by combining EB treatment and bioflocculation (TDS will be removed), make waste water fit for both industrial and agricultural purposes.

Economics of sewage water treatment with e-beam

The conventional treatment of sewage sludge involves stabilization by aeration or digestion (at about 35°C) to reduce the number of pathogens, especially of bacteria such as salmonella by 1 or 2 log units. To obtain a safe hygienic sludge, further reduction of pathogens by about 3 log units is necessary.

Disinfection by chlorination can be problematic in some circumstances. Chlorine can react with naturally occurring organic compounds found in the water supply to produce compounds known as disinfection by-products (DBPs). The most common DBPs are trihalomethanes (THMs) and haloacetic acids (HAAs). Due to the potential carcinogenicity of these compounds, drinking water regulations across the developed world require regular monitoring of the concentration of these compounds in the distribution systems of municipal water systems. The World Health Organization has stated that the “Risks to health from DBPs” are extremely small in comparison with inadequate disinfection (http://www.wikipedia.org).

The design of waste water treatment plant is usually based on the need to reduce organic loads to limit pollution of the environment. Pathogen removal has very rarely been considered as an objective but for reuse of effluents in agriculture, this must be of primary concern. Treatment to remove waste water constituents that may be toxic or harmful to crops, aquatic plants and fish is normally not economically feasible. However, the removal of toxic elements and pathogens that may affect human health needs to be considered (Pereira et al. 2002). Even though our present study is costly, The EB treated sewage water can be effectively used in irrigation as well as in industries as it does not contain any pathogens and some of the physico-chemical parameters are reduced to a safer level (Tables 3, 4).

Conclusion

In searching for cleaner alternative methodologies, the seed treatment technology using ionizing radiation produced by electron accelerators (electron beam—EB) presents is an environmentally friendly alternative without the use of chemicals. In conventional method, the chlorination of water results in the formation of toxic organo-chlorine products, which are mutagenic and carcinogenic in nature. A preliminary study was conducted to evaluate the effects of electron beam radiation on sewage water. Results obtained at dose rate (3 kGy) were shown to be more efficient for the disinfection of sewage water and at same dose rate showed a substantial improvement in waste water quality with an efficient decrease in organic load that lead to a better bioremediation process. On irradiation, the organic matter in sewage water will be degraded via transforming from complex to simpler molecular forms that are easily metabolized by native soil microflora during irrigation. Therefore, this study stresses the advantage of the use of electron beam radiation for sewage water remediation. Similarly at high dose rate (6 kGy) the removal percentage of organic load in sewage increased up to 50% of their initial load. Moreover, our study defends the use of ionizing radiation in two ways for treatment of sewage water: lower dose rate for helping bioremediation and disinfection during tertiary treatment process and higher dose rate for reduction of organic load present in sewage water.

Acknowledgments This work was funded by the DAE (Department of Atomic Energy), BRNS (Board of Research on Nuclear Sciences, BARC (Bhabha Atomic Research Centre), India. So we are giving thanks to DAE, BRNS, BARC and GITAM University for utilising facilities. We (first three authors) would like extend our sincere thanks to Sri. A.S.Khader Scientific Officer D, ILU-6 Electron Beam Facility of the Radiation Technology Development Division of the Bhabha Atomic Research Centre, Mumbai, India, for the constant guidance and all the possible help during the irradiation of sewage samples.
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