RESEARCH ON THE CHANGE OF ECOSYSTEM SERVICES AND THEIR TRADE-OFF IN URBANIZATION AREA

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Abstract. Rapid urbanization causes changes in regional ecosystem services and has an impact on human well-being. Understanding the changes and trade-offs of ecosystem services in urbanized areas is very important to improve the overall benefits of ecosystem services and maintain regional ecological security. Taking the urbanized area between Zhengzhou City and Kaifeng City in Henan Province, China as an example, the changes in ecosystem services and their trade-offs during the “2005-2010-2015” years were analyzed. The results show that in the process of urbanization, the four selected ecosystem services are in a declining state; The relationship between ecosystem services shows obvious dynamic changes. The traditional method of using a single time period or a time node to study trade-offs is often misjudged, and the research on continuous multi-time series should be strengthened in the future.

Keywords: landscape pattern, evaluation, carbon sequestration, habitat quality, landscape aesthetics, grain yield

Introduction

Urbanization brings agglomeration effect, which can effectively promote the comprehensive development of economy, society and culture (Ouyang et al., 2021; Asgarian et al., 2015). At the same time, in the process of urbanization, a large number of natural and semi-natural land has been transformed into construction land, and the land use type has changed significantly (Vahmani et al., 2016). Substance-recycling and energy-flowing of the ecosystem on the surface are strongly disturbed, and the process of the ecosystem cannot be completed, resulting in a serious impact on the service supply of the ecosystem (Smiraglia et al., 2016).

Ecosystem services are all material products and services obtained by human beings from the ecosystem, which are divided into provision services, regulation services, cultural services and support services (Millennium Ecosystem Assessment, 2005; Redhead et al., 2018). Urbanization leads to landscape fragmentation, reduction of natural and semi-natural habitats, etc. A large number of originally synergistic and competitive relationships among ecosystem services are transformed into trade-off relationships, resulting in a continuous decline in the overall supply of ecosystem services in urbanized areas (Ramyar et al., 2020). Urbanization is the most important factor affecting the change of landscape surface pattern in recent decades, and it is also the human behavior that has the greatest impact on ecosystem services (Moein et al., 2018). Ecosystem is facing increasing urbanization pressure. Exploring the impact of urbanization on ecosystem services has become the focus of global change research (Taylor and Hochuli, 2015; Augustynczik and Yousefpour, 2021).

The research on the impact of urbanization on ecosystem services mainly involves the quantitative evaluation (Yang, 2015), spatio-temporal change and simulation (Li et al.,
2020), driving force analysis (Ndong et al., 2020) of ecosystem services and its impact on human well-being (Yang et al., 2021; Xiao et al., 2021). Ecosystem service assessment is the most studied aspect. The initial research methods are equivalent factor method, market value method, shadow engineering method, opportunity cost method, willingness to pay method and so on. With the advancement of research, model evaluation combined with ecosystem service mechanism has gradually become the mainstream (Torres et al., 2021). Common evaluation models include ARIES (Artificial Intelligence for Ecosystem Services) (Villa et al., 2014; Bagstad, 2011), SoLVES (Social Values for Ecosystem Services) (Estoque and Murayama, 2016; Sherrouse et al., 2014) and InVEST (Integrated Valuation of Ecosystem Services and Tradeoffs) (Tallis and Polasky, 2009; Sharp et al., 2014).

The overall research moves from static ecosystem service assessment to dynamic spatio-temporal analysis and scenario simulation (Li et al., 2020). In terms of urbanization, relevant research has moved from the impact theory of urbanization on ecosystem services to the management and regulation of ecosystem services in the process of urbanization (Liang et al., 2021). Direct or indirect adoption of landscape measures (such as ecological process control, scale adjustment, spatial planning, policy intervention, etc.) has become the mainstream of ecosystem service management. However, due to the correlation between most ecosystem services, ecosystem service management without “trade-offs” often leads to another negative ecological result (Xu et al., 2018). In the process of urbanization, the trade-off of ecosystem services has become an important aspect of research (Zheng et al., 2019). However, in the current research, the trade-off determination is mostly concentrated in a single period or a time period, and there is a lack of long-term uninterrupted serial research. Moreover, there is a lack of specific mechanistic explanations between the assessment of ecosystem services and the determination of trade-offs.

Taking the urbanization area between Zhengzhou City and Kaifeng City in Henan Province, China as an example, the changes of four ecosystem services including carbon sequestration services, habitat quality, landscape aesthetics, and grain yield under the background of urbanization in 2005, 2010 and 2015 were analyzed. Then, the trade-off relationships between the four ecosystem services were determined.

Trying to explain the reasons for the trade-off relationship from the mechanism, and discussing the importance of multi-period trade-off research. It is of great significance to strengthen the accuracy of trade-off studies.

Materials and methods

Study area

The study area is the core area of the integrated development of Zhengzhou city and Kaifeng city in Henan Province, China. Since 2015, with the urban development strategy of Zhengzhou to the East and Kaifeng to the west, as well as the implementation of regional plans such as “Zheng Bian Integration” and “Zheng Bian Industrial Belt Planning”, the original agricultural landscape of the study area has been strongly disturbed and ecosystem services have changed greatly.

The study area is located in the east of Zhengzhou and the west of Kaifeng. It starts from Zhengzhou section of Beijing Hong Kong Macao Expressway in the west, Jinming Road in Kaifeng in the East, Lianhuo Expressway in the north and national highway 310 in the south. The geographical location is between 34°72'- 34°85′N and 113°81'- 114°30′E, with a total area of about 47314 hm². The details are shown in Figure 1.
Data sources

The image data are come from the geospatial cloud data platform. Three time periods of the study area in May 2005, 2010 and 2015 were obtained and processed. On the basis of the previous investigation, the landscape elements in the study area are divided into five types (farmland, forest land, water area, construction land, and unused land) by visual interpretation. The image data of the study area were processed by GIS 10.0 to obtain the landscape pattern data of the study area in three periods. According to the previous research results (Fan et al., 2018), it is determined that the best analysis grain size of landscape pattern in the study area is 20 m × 20 m.

The grain yield and area data in Henan Province mainly comes from the Henan Provincial Statistical Yearbook, National Agricultural Product Cost and Benefit Data Compilation, Zhongmu County Statistical Yearbook and other related materials.

Data related to carbon storage are mainly from local forestry departments. Part of the data comes from previous scholars’ research in this region, and has been revised based on the climatic conditions of the study year (Huang et al., 2014, 2006; Guang, 2007; Chen et al., 2007; Peng et al., 2013).

Ecosystem services assessment methods

Carbon sequestration service

Carbon sequestration services are evaluated using the Carbon Model module in the InVEST model. The formula is shown in Equation 1:
where \( C_{\text{stored}} \) is the total carbon storage; \( C_{\text{above}} \) is the above-ground material carbon storage (t/hm²); \( C_{\text{below}} \) is the underground material carbon storage (t/hm²); \( C_{\text{soil}} \) is the soil carbon storage (t/hm²); \( C_{\text{dead}} \) is the litter carbon storage (t/hm²).

### Habitat quality assessment

The Habitat Quality module in the InVEST model was used to evaluate the habitat quality of the study area. The formula is as follows (Eq. 2):

\[
Q_{xj} = H_j - H_j \times \frac{D_{xj}^2}{D_{xj}^2 + K}
\]

where \( Q_{xj} \) is the habitat quality of grid X in land use type J; \( D_{xj} \) is the stress level of grid X in land use type J; K is the semi-saturation constant, usually half of the maximum value of \( D_{xj} \); Hj is the habitat suitability of habitat J in land use type. When habitat is studied only from a macro perspective without involving species, the Hj value is often defined by dichotomy; Z is the normalization constant, generally 2.5. The dxj calculation is shown in Equation 3:

\[
D_{xj} = \sum_{r=1}^{R} \sum_{y=1}^{Y} \left( \frac{W_r}{W_j} \right) i_{rxy} \beta_x S_{jr}
\]

In the formula, R is the habitat stress factor; y is the grid number of the habitat stress factor r grid layer; Yr is the number of grids occupied by the habitat stress factor; Wr is the weight of the stress factor, indicating that a certain stress factor affects all habitats. The relative damage degree of r is 0-1; ry is the stress factor value of grid y (0 or 1); irxy is the stress factor value of grid y and the stress level of ry to habitat grid x; \( \beta_x \) is accessibility level of grid x, and the value 0-1. 1 means extremely easy to reach; Sjr is the sensitivity of the habitat type j to the stress factor r, and the value is 0-1, the closer the value is to 1, the more sensitive it is. irxy can be obtained by Equation 4:

\[
i_{rxy} = 1 - \frac{d_{xy}}{d_{max}}
\]

In the formula, dxy is the straight-line distance between grid x and grid y; dmax is the maximum influence distance of stress factor r.

### Landscape aesthetics

Landscape aesthetics is evaluated by the expert knowledge-based ecosystem service value assessment method constructed by Xie et al. (2015). This is a dynamic equivalent factor method for the evaluation of ecosystem services in China, which has been verified and recognized in many regions of China (Fan, 2016). The equivalent factor method is a dynamic equivalent data accounting based on the sowing area and net profit of local main agricultural products in the current year. The calculation is shown in Equation 5:

\[
D = S_t \times F_t + S_n \times F_n + S_s \times F_s
\]
where $d$ represents the ecosystem service value of one standard equivalent (yuan/hm²); $S_r$, $S_w$ and $S_c$ respectively represent the percentage of the sown area of rice, wheat and corn in the total sown area of the three crops in the study year; $F_r$, $F_w$ and $F_c$ respectively represent the national average net profit per unit area of rice, wheat and corn in the study year (yuan/hm²).

After the $D$ value is obtained, combined with different land use types, the value equivalent of each land use type can be calculated. Then multiply the area of the land use type to get the landscape aesthetic value of the land use type.

**Grain yield**

It is calculated based on the unit area yield and planting area of grain crops in the study year.

**Result**

**Carbon sequestration services**

The four carbon pool values and landscape pattern data in the study area were input into the InVEST model to calculate the carbon sequestration service distribution map in three periods (Fig. 2).

![Image of carbon storage distribution](image-url)
Subsequently, using the spatial statistical tool of ArcGIS 10.0, the carbon storage in the three periods of the study area was summarized and mapped in Figure 3.

![Figure 3. Changes in carbon storage in the study area from 2005 to 2015](image)

The overall carbon storage in the study area is in a downward trend, from 2005 to 2015, it decreased by about $3.9 \times 10^8$ kg. In 2010, it fluctuated slightly, reaching $4.52 \times 10^9$ kg, which was $3 \times 10^7$ kg higher than that in 2005. The carbon storage in the study area is in a trend of rising first and then falling (↑, ↓).

**Habitat quality assessment**

The 20 m × 20 m landscape pattern data, threat factor layer data, threat factor attribute table and the sensitivity attribute table data of different landscape types to threat factors in the three periods of the study area are input into the Habitat Quality module in the InVEST model for calculation, and evaluation results of habitat quality is obtained. Then, the habitat quality of the three periods of 2005, 2010 and 2015 is divided into three intervals from low to high under the GIS platform: poor (0-0.2), medium (0.21-0.55), and good (0.56-0.9). The details are shown in Figure 4.

The closer the habitat value is to 1, the higher the habitat quality, and the closer the habitat value is to 0, the worse the habitat quality. It can be seen from the above three figures that the areas with poor habitats in the study area are generally dominated by construction land, while the areas with good habitats are dominated by forest land and water areas, and the habitat quality of farmland is between construction land and forest land. The area proportion comparisons of different habitat qualities in the study area in 2005, 2010 and 2015 are shown in Figure 5. According to Figure 5, it can be concluded that the overall habitat quality in 2005 > the overall habitat quality in 2010 > the overall habitat quality in 2015. Therefore, the change trend of habitat quality is (↓, ↓).

**Landscape aesthetic evaluation**

According to the method described above and the relevant data of the study area, the aesthetic value of the study area in the three periods was obtained (Table 1). According to Table 1, the landscape aesthetic value of the study area increases first and then decreases (↑, ↓).
Grain yield evaluation

According to the data of crop yield and area in the study area, the grain production value in the study area is obtained (Table 2). According to Table 2, it can be concluded that the grain yield in the study area is in a downward trend (↓, ↓).

![Habitat quality of the study area in 2005](image1)

![Habitat quality of the study area in 2010](image2)

![Habitat quality of the study area in 2015](image3)

Figure 4. Distribution of habitat quality in the study area in 2005, 2010 and 2015

![Proportion of different habitat quality areas in 2005 and 2010](image4)

![Proportion of different habitat quality areas in 2010 and 2015](image5)

Figure 5. Comparison of different habitat quality areas in different periods

| Year | 2005  | 2010  | 2015  |
|------|-------|-------|-------|
| Total| $3.79 \times 10^7$ | $4.25 \times 10^7$ | $2.42 \times 10^7$ |

Table 1. Aesthetic value of the study area in different periods (RMB: yuan)
Table 2. The grain production value of the three periods in the study area

| Year | 2005 | 2010 | 2015 |
|------|------|------|------|
| Total output (kg) | 1.91×10^8 | 1.88×10^8 | 1.69×10^8 |

Overall trade-off relationship of four ecosystem services

Combined with 3.1-3.4, the change trend of four types of ecosystem services is tabulated (Table 3).

Table 3. Change trend of four types of ecosystem services

| Types of ecosystem services | Trend from 2005 to 2010 | Trend from 2010 to 2015 |
|-----------------------------|--------------------------|------------------------|
| Carbon sequestration services | ↑                        | ↓                      |
| Grain production            | ↓                        | ↓                      |
| Habitat quality             | ↓                        | ↓                      |
| Landscape aesthetics        | ↑                        | ↓                      |

As shown in Table 3, the relationship between any two ecosystem services in the same period can be obtained. But the relationship between them may be inconsistent at different periods. For example, carbon sequestration services and food production showed a trade-off relationship from 2005 to 2010, but a synergistic relationship from 2010 to 2015. Therefore, studying the relationship of ecosystem services in a short period is likely to lead to misjudgment results. Looking at the entire time period 2005-2020 and 2010-2015, carbon sequestration services (↑, ↓) and aesthetic landscape (↑, ↓) show a synergistic relationship, as well as the relationship between food production (↓, ↓) and habitat quality (↓, ↓).

Discussion

Research period of ecosystem service trade-offs

At present, the research on ecosystem service trade-off is mainly based on spatial distribution mapping and correlation analysis (Deng et al., 2020), and the research is mostly concentrated in a single period or between two periods. There are few studies on three or more consecutive periods (Yin et al., 2019; Wang et al., 2020). This study found that the trade-off study in a single period may misjudge the relationship between ecosystem services. Therefore, strengthening multi period research is helpful to improve the scientificity of research.

Determination of research methods on trade-offs

There are many previous studies on trade-offs. Almost all studies first evaluate ecosystem services, and then make trade-offs according to the changes of results (Li et al., 2021). There are many evaluation methods of ecosystem services. Different evaluation methods often get different evaluation results of ecosystem services, so the accuracy of trade-off judgment is also affected. In addition, the traditional research does not analyze the relationship between different ecosystem service assessment methods, which is also the limitation of the current research.
Take the carbon sequestration service and landscape aesthetics studied in this paper as an example. The carbon module in the InVEST model is used for valuing carbon sequestration service. The calculation method of carbon reserves in the study area can be simplified as: area of different landscape elements \( \times \) Carbon reserves per unit area, and then sum; The aesthetic landscape value is evaluated by dynamic equivalent method, which can also be simplified as: the area of different landscape elements \( \times \) The value equivalent per unit area is then summed.

These two ecosystem services have a relatively consistent calculation method: the area of different landscape elements \( \times \) the ecological value service per unit area, and then summed. In terms of calculation principle, these two ecosystem services have relatively consistent calculation methods. At the same time, in the process of urbanization, the change area of landscape elements they contain is also the same, so carbon sequestration services and landscape aesthetics show a synergistic relationship.

In fact, the ecosystem service assessment method is the determinant of the trade-off relationship and the basis for explaining the trade-off of ecosystem services. Therefore, a scientific method of ecosystem services assessment is a prerequisite for trade-off studies.

**Conclusion**

Through the research on the changes of ecosystem services in the process of urbanization in the study area in 2005-2010-2015, some obvious conclusions and suggestions for future research are as follows:

**Main conclusions**

*Ecosystem services are greatly affected by urbanization*

The four ecosystem services in this study are selected from supply service, support service, regulation service and cultural service. They have all undergone significant changes in the process of urbanization and are obviously representative. In addition, a large number of ecosystem services exist on the surface landscape as a carrier, and urbanization drives changes in landscape patterns, which will inevitably lead to a large number of changes in ecosystem services.

*Ecosystem services research only in a single period may misjudge the trade-off relationship between them*

For example, in this study, carbon sequestration services and food production showed a trade-off relationship in 2005-2010, and a synergistic relationship in 2010-2015. Analysis of a single time period will misjudge their relationship.

*The scientificity of ecosystem services assessment directly affects the accuracy of trade-offs*

From the trade-off study of four ecosystem services in this paper, it can be seen that the evaluation method of ecosystem services has obvious explanatory power for the trade-off judgment between them. Inconsistent ecosystem service assessment methods may lead to different trade-off results. Therefore, a scientific ecosystem service assessment method involving mechanism is very important for trade-off research.
Main recommendations

Ecosystem services research needs to choose research methods derived from ecosystem mechanisms

Evaluating ecosystem service trade-offs requires quantifying different ecosystem services. However, there are many evaluation methods for the same ecosystem service, and the evaluation methods for different ecosystem services are also different. Different methods will lead to different results. Therefore, scientific evaluation methods which were selected from the ecosystem mechanism, can ensure the accuracy of the research.

Accurate trade-off law research needs continuous and long-time data support

Due to the dynamicity of ecosystem services, the trade-offs between them will inevitably change. Short time trade-off performance cannot represent the accurate trade-off law. In future research, it is necessary to add long-time series data to ensure the accuracy of research results.

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