Diverse traits of aquatic plants cannot individually explain their consumption by the generalist gastropod *Biomphalairia glabrata*

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ABSTRACT

Several experimental studies on aquatic plants have reported the prevalence of chemical defense mechanism against herbivory, as opposed to structural, life-forms or other traits. Here, our laboratory feeding experiments and integrative analysis explored the relationship among palatability (fresh or reconstituted plants used as artificial diet) and various chemical/nutritional traits (*i.e.*, contents of dry mass, ash, nitrogen, protein, and phenols) of diverse aquatic plants and their susceptibility to consumption by the generalist gastropod *Biomphalaria glabrata*. *Biomphalaria glabrata* consumed all of the assayed aquatic plants in a hierarchical yet generalized way, with the consumption of fresh plants, their reconstituted forms and defensive properties of lipophilic extracts not being significantly correlated with plant physical or chemical traits to determine the feeding preference of the gastropod. Our results do not reveal a prevalence for a specific plant attribute contributing to herbivory. Instead, they indicate that the susceptibility of aquatic plants to generalist consumers is probably related to a combination of their chemical and physical properties, resulting in moderate grazing rates by generalist consumers.

Subjects Ecology, Plant Science, Zoology, Freshwater Biology

Keywords Herbivory, Aquatic biology, Macrophytes, Chemical defense, Physical defense, Structural defense, Secondary metabolites, Plant life-form, Plant defense, Susceptibility to herbivory

INTRODUCTION

Herbivory has long been recognized as a selective and structuring component in terrestrial (*Hastin, Smith & Slobodkin, 1960*) and marine environments (*Lubchenco, 1978*). Although recognized later, herbivory on aquatic plants is also considered selective and intense (*Newman, 1991; Lodge et al., 1998*) and it is capable of removing about 44–48% of plant biomass, exerting cascade effects on community structure in aquatic environments.
Terrestrial and aquatic environments share several structural features, but the ecology and evolution of chemical defense and other components of plant-herbivore interactions remain poorly understood in freshwater environments.

Aquatic plants are known to produce a wide variety of secondary metabolites, such as phenols, terpenes, alkaloids, polyketides, among others (Ramesh, Rajan & Santhanan, 2014). Give the selective pressure of herbivory (Lodge, 1991) and the considerable diversity of defensive metabolites (Ramesh, Rajan & Santhanan, 2014), have compounds likely play important ecological roles in freshwater environments. Indeed, diverse studies, from simple defensive chemical content analysis (e.g., Ostrofsky & Zettler, 1986) to more complex experimental studies (e.g., Bolser & Hay, 1998) support that likelihood.

In one of the first such experimental studies, extracts of aquatic plant species were shown to significantly reduce herbivory (Bolser et al., 1998). Extracts of two species of aquatic plants, *Nuphar luteum* var. *macrophyllum* and *Alternanthera philoxeroides*, were found to inhibit the consumption by the omnivorous crayfish *Procambarus clarkii*, though other aspects, such as structure and nutritional content also contributed to the consumer’s food selectivity (Cronin et al., 2002). Defensive chemicals in the aquatic plant species *Micranthemum umbrosum* were shown to inhibit consumption by the crustaceans *Procambarus spiculifer* and *P. acutus*, as well as the fish *Ctenopharyngodon idella* and these chemically-mediated relationships are structuring for the freshwater plant community (Parker et al., 2006).

In the context of allelopathic defense, chemicals in the exudates of *Eichhornia crassipes* and *Salvinia biloba* repel the cladocera *Ceriodaphnia dubia* and decrease its life expectancy, increasing its age at first reproduction and lowering the number of eggs generated per female (Gutierrez & Paggi, 2014). Chemical exudates from aquatic plants also repel some planktonic organisms (Pennak, 1973; Dorgelo & Heykoop, 1985).

Chemical defense by aquatic plants can also be dynamic or induced upon perceiving a cue that indicates an increased risk of the plant being grazed. The specialist beetle *Galerucella nymphaeae* induces chemical defense in the water lily *N. luteum* var. *macrophyllum* that inhibits consumption by the generalist crayfish *P. clarkii* (Bolser & Hay, 1998). Chemical defense in *Myriophyllum spicatum* induced by the activity of the insect *Acentria ephemerella* promotes protection of the apical regions of this plant, preventing future attacks by this same species (Fornoff & Gross, 2014). In *Cabomba caroliniana*, chemical defense is induced by *P. clarkii* and by the snail *Pomacea canaliculate*, which reduces this plant’s palatability of this plant for both consumers and also suppresses the growth of co-occurring microbes (Morrison & Hay, 2011). Rapid production of nitrogen in apical regions of *M. spicatum* upon consumption by the herbivorous insect *Acentria ephemerella* supports the existence of a nitrogen-based chemical defense system in these young plant parts (Rothhaupt, Fornoff & Yohannes, 2015).

A review highlighted that the chemical defenses in aquatic plants are extensive but remain largely unexplored (Gross & Baker, 2012). There is little quantitative information on what determines the susceptibility of aquatic plants to herbivory or the basis of choice by herbivores (Lodge, 1991; Newman, 1991). However, several experimental studies support
an apparent prevalence of chemical defense over other plant traits in herbivore feeding preferences (e.g., Kubanek et al., 2001; Gutierrez & Paggi, 2014).

Although various structural plant defenses, such as spines, thorns, or pubescence/hairiness are rare or even non-existent in aquatic plants (Vermeij, 2016), few structural aspects of such plants have been implicated in the feeding preferences of freshwater herbivores (Chambers, Hanson & Prepas, 1991; Cronin, 1998). Instead, plant nutritional value (Fraser, Chavez & Paloheimo, 1984; Doucet & Fryxell, 1993; Kreider & Watts, 1998) and presence of chemical defense are widely considered the main traits determining aquatic plants palatability (Cronin et al., 2002; Sotka et al., 2009). However, there is still a lack of integrated studies exploring all aspects contributing to plant-herbivore interactions in freshwater environments and that can elucidate the ecology and evolution of these trophic relationships.

Here we deployed an integrated experimental approach to investigate the relationship between the susceptibility of diverse aquatic plant life-forms, from emergent and floating to fully submerged species, and several of their physical and chemical traits in terms of their consumption by the generalist gastropod Biomphalaria glabrata.

**MATERIALS & METHODS**

**Field sites and sample collections**

Thirteen aquatic plant species belonging of different families and that represented a wide range of life-forms (according Cook, 1996) and taxonomic groups were target (Table 1). All plants were collected in aquatic habitats inside the Restinga de Jurubatiba National Park, Rio de Janeiro State, Brazil (SISBio 30299). Immediately after collection, the plants were conditioned in thermal boxes and transported to the laboratory. The species were identified by one of the authors (Tatiana U.P. Konno) and voucher specimens have been deposited at the Herbarium of the Rio de Janeiro Federal University, Brazil. Samples of vegetative parts of each species were separated into three groups: (1) fresh plants for use in susceptibility to consumption assays (within 48 h of collection); (2) lyophilized and ground for the assays using reconstituted plants (as artificial food) and the nutritional analyses; (3) chemical extractions to investigate the defensive capabilities.

Specimens of the generalist herbivorous gastropod B. glabrata (Thomas, Nwanko & Sterry, 1985) (shell diameter = 1.7 ± 0.3 cm) were collected from Cabiúnas lagoon (22°16’S, 41°41’W) by actively searching through clumps of aquatic plant using a sieve. Upon collection, the gastropods were conditioned in thermal aerated boxes containing freshwater and transported to the laboratory. There, they were maintained in plastic containers (10 L) with freshwater changed daily, constant aeration, constant temperature (22 °C) and a 12:12 h light:dark photoperiod. They were fed on lettuce for 48 h until 1 h before the beginning of experiments to avoid starvation that could promote unnaturally feeding behavior during our assays (Cronin & Hay, 1996).

**Susceptibility of the fresh whole aquatic plants to B. glabrata**

The susceptibility of nine species of aquatic plants to consumption by B. glabrata was assessed according to a previous study (Cronin et al., 2002). Each experimental unit
Table 1  Information on aquatic plant species used in the assays, including family, habit (life-form), collected site coordinates, and voucher number.

| Family            | Species                              | Habit (life-form)   | Collection site | Voucher   |
|-------------------|--------------------------------------|---------------------|-----------------|-----------|
| Alismataceae      | Hydrocleys nymphoides (Willd.) Buchenau | Hydrophyte: emergent free floating | 22°17'0" S 41°41'1" W | RFA39121 |
| Araceae           | Pistia stratiotes L.                 | Hydrophyte: emergent free floating | 22°17'0" S 41°42'2" W | RFA39122 |
| Cyperaceae        | Eleocharis interstincta (Vahl) Roem. & Schult. | Hydrophyte: emergent | 22°17'0" S 41°41'1" W | RFA39123 |
| Lemnaceae         | Utricularia foliosa L.               | Hydrophyte: submerged | 22°17'0" S 41°41'1" W | RFA39244 |
| Menyanthaceae     | Nymphoides humboldtiana (Kunth) Kuntze | Hydrophyte: emergent floating | 22°12'0" S 41°29'1" W | RFA39117 |
| Nymphaeaceae      | Nymphaea pulchella (Salisb.) DC.      | Hydrophyte: emergent floating | 22°17'0" S 41°41'1" W | RFA39245 |
| Pontederiaceae    | Eichhornia azurea (Sw.) Kunth         | Hydrophyte: emergent floating | 22°17'0" S 41°41'1" W | RFA39120 |
| Potamogetonaceae  | Potamogeton illinoensis A. Benn.      | Hydrophyte: emergent floating | 22°16'0" S 41°39'1" W | RFA41421 |
| Pteridaceae       | Ceratopteris thalictroides (L.) Brongn. | Hydrophyte: emergent | 22°17'0" S 41°41'1" W | RFA39211 |
| Rubiaceae         | Oldenlandia salzmannii (DC.) Benth. & Hook. f. ex B.D. Jacks. | Hydrophyte: submerged to Helophyte | 22°16'0" S 41°39'1" W | RFA39119 |
| Salviniaeae       | Salvinia auriculata Aubl.             | Hydrophyte: emergent free floating | 22°16'0" S 41°39'1" W | RFA39242 |
| Typhaceae         | Typha domingensis Pers.               | Fixed emergent       | 22°17'0" S 41°41'1" W | RFA39234 |
| Xyridaceae        | Xyris brevifolia Michx.               | Helophyte: submerged  | 22°10'0" S 41°31'1" W | RFA39240 |

Notes.
*Life-forms according to Cook (1996).*

(replicate) consisted of a transparent plastic container (1 L) supplied with non-recirculating freshwater, in which one sample (approximately equal volumes to provide comparable encounter rates) of each of the nine types of aquatic plant was simultaneously offered to a single individual of B. glabrata for 3 days (n = 14) to determine the differential susceptibility of them to this gastropod. Only vegetative parts of each of the nine plant species were used. One sample of each plant species was set up under identical conditions, but without B. glabrata (as a control) to verify possible autogenic changes in biomass (independent of herbivory). Each plant sample was weighed at the beginning and at the end of the experiment, and the biomass consumed by B. glabrata was expressed as a percentage (%) of consumption, calculated according to the equation [H₀ × (C_f/C₀) − H_f] × 100; where H₀ and H_f represent the biomass of the plant sample exposed to herbivory before and after the assay, respectively, and C₀ and C_f represent the biomass of corresponding control samples of the species from before and after the assay. For this bioassay, 14 replicates of each plant species were used.

Consumption of reconstituted aquatic plants by B. glabrata
To determine if plant structure influenced the consumption by B. glabrata, samples of the 13 collected aquatic plant species were frozen, freeze-dried, ground into a fine powder, and stored at −4 °C for later preparation as artificial foods. Artificial foods (i.e., reconstituted plants) were prepared in agar gel to reproduce the natural dry mass concentration of each species, as described previously (Bolser & Hay, 1998; Cronin et al., 2002). The amount of powdered plant incorporated as artificial foods varied for each species due to the specific relationships between dry mass and wet mass. The artificial food comprising each plant species was prepared by mixing 5.0 mL of distilled water with 0.36 g of agar, heating the mixture in a microwave oven until boiling, and then pouring it into cold water containing...
desired amounts of freeze-dried powder for each species (i.e., each species was rehydrated with the same amount of cold water to reestablish its wet mass). This mixture was allowed harden onto a plastic screen mesh (∼1 mm² per aperture) and flattened between two layers of wax paper. After cooling and solidification, the screen mesh was cut into smaller pieces of 10 × 10 mm² (total of 100 squares). Accordingly, all artificial foods presented the same morphology and hardness, but the original nutritional value and taste of each species was maintained as much as possible. For this bioassay, 30 replicates of each plant species were used.

The 13 artificial foods were simultaneously offered to one specimen of *B. glabrata* in 1.0 L plastic containers supplied with non-recirculating freshwater (*n* = 30) and consumption of them was calculated by scoring the number of squares of artificial food that were consumed. A fixed consumption stopping rule was applied (*Lockwood III, 1998*), where all assays were terminated once >50% of all artificial food had been consumed.

**Defensive capacity of the plant extracts against *B. glabrata***

The defensive capabilities of the extracts from ten aquatic plants against *B. glabrata* were evaluated by assaying *B. glabrata*’s consumption of an artificial food containing the extract (treatment) and a respective control (without extract). A lipophilic extract of each plant species was incorporated at a naturally occurring concentration (extract weight equivalent to the plant dry weight) to artificial food (treatment and control) prepared as described previously (*Bolser & Hay, 1998; Cronin et al., 2002*). For treatment foods, lipophilic aquatic plant extracts were first dissolved in CH₂Cl₂, added to a freeze-dried lettuce:broccoli mixture (1:1) that lacks defensive compounds, and then the solvent was removed by rotary evaporation. This procedure resulted in a uniform coating of the lipophilic extracts on the lettuce:broccoli plant particles (*Hay, Kappel & Fenical, 1994*). For the control food, only the solvent was incorporated into the lettuce:broccoli mixture. Artificial foods (treatment or control) were then prepared by adding 0.36 g of agar to 5.0 mL of distilled water, heating the mixture in a microwave oven until boiling, and then pouring it into 2 mL of cold water containing 0.5 g of freeze-dried powder of the lettuce:broccoli (1:1) mixture with (treatment) or without (control) extract. Treatment and control artificial foods were hardened separately onto a plastic screen mesh (∼1 mm² per aperture) and flattened between two layers of wax paper. After the artificial foods had cooled and solidified, the screen mesh was cut into strips comprising 100 squares (10 mm²) of artificial food pieces of 10 × 10 mm square (control and treatment). The feeding assays involved simultaneously offering a treatment (one particular plant extract) and a control strip of artificial food to one individual *B. glabrata* as a two-way choice. The assays were carried out in small plastic aquaria (500 mL) containing 250 mL of the water collected from Cabiúnas Lagoon. The number of replicates for each extract (species) varied between 18 and 26, and gastropod specimens were not used for more than one replicate.

Food consumption was estimated as described above under the section *Consumption of reconstituted freshwater plants by B. glabrata*, and the fixed-consumption stopping rule was again applied. Percentage inhibited consumption was was calculated according to the differences between consumption of control and treatment artificial foods, quantified as
% Inhibition = [(C – T)/C] × 100 for each replicate, where C and T correspond to control and treatment squares consumed, respectively.

**Extracts preparation**
Extracts were obtained from a total of 10 aquatic plant species. Their vegetative parts were separated and air-dried under shade and in a climatized room at ambient temperature (mean 21 °C) for 48 h. In order to minimize the intra-individual variability among organisms, each plant was gathered and subjected to extraction together. Dried plants (X. brevifolia, 47.7 g of dry weight, d.w.; E. interstincta 27.3 g d.w.; T. domingensis 26.3 g d.w.; P. illinoensis 19.6 g d.w.; N. humboldtiana 95.0 g d.w.; E. azurea 54.7 g d.w.; S. auriculata 31.3 g d.w.; P. stratiotes 42.0 g d.w.; O. salzmannii 29.8 g d.w.; U. foliosa 193.0 g d.w.) were ground down and extracted with a mixture of dichloromethane:methanol, 1:1 (HPLC grade, Tedia), using 20 mL of solvent for each 1 g of plant dry mass for one day at room temperature. Then the solvent was filtered out and the biological material was re-extracted using the same solvent/plant mass ratio. This procedure was repeated twice, and then the solvents removed from these extractions were combined and dried by rotary evaporation (Buchi R210, V850, B491, Switzerland) to yield a total of 10 different plant extracts that were used in the feeding assays and to quantify phenolic contents. The extract from *Potamogeton illinoensis* was further subjected to chemical profile analysis by Proton Nuclear Magnetic Resonance (¹H NMR).

**Physical, nutritional and chemical properties of the aquatic plants**
Several tissue properties known to influence herbivory were measured for the 13 collected aquatic plant species. Powdered aquatic plants were analyzed for total nitrogen using the Kjeldahl method, as described previously (Conklin-Brittain et al., 1999), and the protein content was estimated by multiplying the value for total nitrogen by a factor of 6.25 (see Barbarino & Lourenço, 2005). Ash content was accurately determined via combustion of powdered biomass in a muffle oven at 550 °C (Jones, 2001). Lipids were extracted with diethyl ether using a Soxhlet system (see details in Ramluckana, Moodley & Bux, 2014). Results for ash and lipids contents are expressed as a percentage (%) by dry mass, total nitrogen as g% nitrogen (N) and protein content as g% protein (P). Phenol content was quantified from crude extracts of the plants using a modified Folin-Ciocalteu colorimetric method (Zhang et al., 2006) and the results are expressed as milligrams of GAE (gallic acid equivalent)/g of the extract. All analyses were performed in triplicate.

**Chemical profile of *Potamogeton illinoensis* by ¹H NMR**
The chemical profile *P. illinoensis*, the most active extract against *B. glabrata*, was analyzed by ¹H NMR in order to identify its major lipophilic compounds. NMR analysis was performed on the Varian VNMRs apparatus with 11.75 T (500 MHz) magnet. Deuterated chloroform was used to solubilize the sample.

**Data analysis**
To evaluate the various relationships between the consumption by *B. glabrata* and the different plant traits, data on consumption and the various physical, nutritional and
chemical properties of the plant species were compared according to species and also plant life-forms.

Non-parametric tests were used to analyze the herbivory data, as data was not show normally distributed. To compare among species, data on percentage consumption (%) whole plants and artificial foods (reconstituted plants) by *B. glabrata* were analyzed using the non-parametric Friedman test due data dependency (all species were offered simultaneously to the herbivore). Therefore, the comparison between the different life-form groups the % of consumption data of the whole and reconstituted aquatic plants were analyzed by Kruskal-Wallis non-parametric analysis of variance due to inequality in the sample size of the groups. For pair-wise comparisons of the non-parametric tests, the post-hoc Durbin-Conover and Dwass-Steel-Critchlow-Fligner test were performed to the Friedman and Kruskal-Wallis analysis, respectively. One Way ANOVA, followed by a post-hoc Tukey test was applied for comparisons of physical, nutritional and chemical properties of the aquatic plants. Pearson correlation coefficients were used to evaluate relationships among percentage consumption of whole plants, artificial reconstituted plants, extract-mediated inhibition of consumption, and various physical, nutritional and chemical traits of the plant species. For all statistical analyses, results were considered significant when \( p < 0.05 \). All statistical analyses of variance and correlation were performed in Jamovi version 1.6 freeware (*R Core Team. 2020; Jamovi The Project, 2021*).

A multivariate approach was conducted using the exploratory descriptive Principal Components Analysis (PCA) to summarize all data and allow access and visualization of a possible relationship between plant traits and herbivory inhibition. The consumption data, and physical, nutritional and chemical traits were summarized in a matrix that was submitted to a PCA. The physical, nutritional and chemical traits were used as supplementary quantitative variables. The PCA was performed in R version 4.0.3 with the function “PCAshiny” in the “Factoshiny” package (*Vaisie, Monge & Husson, 2021*). The scores results were represented by scatter plots of the principal components (PC) for visualization of the discriminant pattern of the samples (species and plant life-form) according to their distribution along the first two PC axes (PC1 =Dim 1 and PC2 =Dim 2). The loading results (consumption data, physical, nutritional and chemical traits) used to discrimination of the samples were represented by scatter plots.

**RESULTS**

**Susceptibility of intact plants to *B. glabrata* herbivory**

The percentage of fresh biomass of the nine aquatic species consumed by *B. glabrata* ranged from 2.08% (*E. interstincta*) to 0.45% (*P. illinoensis*) (median values, Fig. 1A), revealing three groups, two of which differed significantly from each other \( (\chi^2 = 34.15, df = 8, p < 0.001, \text{Tukey test}) \). All plant species were consumed to some degree, but with the following hierarchy in terms of diminishing preference (Fig. 1A): *E. interstincta > N. pulchella > E. azurea > H. nymphoides (group 1) > O. salzmannii > T. domingensis > N. humboldtiana (intermediate group, not significantly different from groups 1 and 2, \( p > 0.05 \) > S. auriculata > P. illinoensis (group 2).
There was no significant pattern in terms of plant biomass consumed according to plant life-forms ($H = 349.94$, $df = 4$, $p = 0.055$) (Fig. 2A), though *B. glabrata* had consumed more emergent plants (2.08%) than emergent floating plants (0.90%).

**Consumption of reconstituted aquatic plants by *B. glabrata***

When we eliminated interspecific differences arising from physical/structural characteristics, but retained the taste, nutritional and chemical properties of the thirteen aquatic plants by offering reconstituted plants as artificial food to *B. glabrata*, we obtained different results from those observed for fresh plants (Fig. 1A). Again, the gastropod consumed all 13 plants (Fig. 1B), but in significantly different amounts ($\chi^2 = 144.36$, $df = 12$, $p < 0.001$, Tukey test). Moreover, distinct groupings were not observed and the hierarchy of diminishing preference was quite different: *T. domingensis* > *X. brevifolia* > *N. pulchella* > *E. interstincta* > *O. salzmannii* > *H. nymphoides* > *C. thalictroides* > *P. stratiotes* > *N. humboldtiana* > *S. auriculata* > *E. azurea* > *U. foliosa* > *P. illinoensis*. For instance, *Eichornia azurea* that was highly consumed as a fresh plant was one of the least...
Figure 2  Results of the choice feeding assays results with of *B. glabrata*, analyzed according to the plant life-forms: (A) Susceptibility of nine fresh whole plants; (B) consumption of thirteen reconstituted aquatic plants; and (C) defensive effect of ten plant extracts. The same letters on bars indicate the treatments are not significantly different. Lower and upper box boundaries 25th and 75th percentiles, respectively, line inside box median and lower and upper error lines 10th and 90th percentiles, respectively. *H* = Helophyte, *FE* = Fixed emergent, *E* = Emergent, *EF* = Emergent Free Floating, *S* = Submerged, *SH* = Submerged to Helophyte.

preferred as a reconstituted food. However, *P. illinoensis* remained unpalatable the least palatable (3.00% consumption, median value), suggesting it harbor a component that exerts an inhibitory effect on consumption by *B. glabrata*. In contrast, *T. domingensis* that was one of the least susceptible to consumption as a fresh plant, was the most preferred (%
consumption, median value) as reconstituted food, indicating that it possesses structural components that deter consumption by this gastropod.

Considering the life-forms of the plants reconstituted as artificial food (Fig. 2B), we noted a significant decreasing preference by *B. glabrata* (*H* = 197.10, *df* = 6, *p* < 0.001, Tukey test): fixed emergent > helophyte > submerged to helophyte > emergent > emergent free floating > emergent floating > submerged. Accordingly, *B. glabrata* appears to exhibit a preference for consuming aquatic plants that occupy the marginal areas of the lagoons from which they were collected.

**Defensive capabilities of the plant extracts**

The nine extracts of aquatic plants we assayed exhibited significantly different defensive effects (*H* = 101.10, *df* = 8, *p* < 0.001, Tukey test) against consumption by *B. glabrata* (Fig. 1C), with inhibition (%) ranging from 100% (*P. illinoensis*) to 13.3% (*S. auriculata*), and with a hierarchy of diminished inhibition of: *P. illinoensis* > *X. brevifolia* > *T. domingensis* > *E. azurea* > *N. humboldtiana* > *U. foliosa* > *E. interstincta* > *P. stratiotes* > *S. auriculata*. In contrast, the extract from *O. salzmannii* significantly stimulated consumption by the gastropod. When we focused in plant life-forms, we found significant differences between species in terms of extract-mediated herbivory inhibition (*H* = 93.00, *df* = 6, *p* < 0.001, Tukey test), with the following hierarchy (Fig. 2C): emergent floating = helophyte > fixed emergent > submerged > emergent > emergent free floating. Chemicals from submerged and helophyte plants stimulated the consumption. For the emergent floating plant *P. illinoensis*, this result reiterates that obtained in previous assays about susceptibility to consumption by *B. glabrata* (Figs. 1A and 2A), as well as their consumption as reconstituted food (Figs. 1B and 2B), since in both they were the least consumed food items.

**Physical, nutritional and chemical properties of the aquatic plants**

The physical, nutritional and general chemical defense properties we assessed varied significantly among the plant species and also when analysed to life-forms (Table 2) Dry mass, ash, lipid, total nitrogen, protein, and phenol levels varied from five-fold (g% nitrogen) to almost thirty-fold (% lipid). *Salvinia auriculata* displayed the lowest percentage dry mass, whereas *X. brevifolia* scored highest in this respect, reflecting their life-form as submerged and helophyte plants, respectively. Ash (%) was lowest in *E. interstincta* (7.10 ± 0.18%) and highest in *P. stratiotes* (39.42 ± 0.47%), and more generally was lower among emergent free-floating plants. Lipid contents displayed the widest variation, ranging from 1.16 ± 0.07% in *E. interstincta* to 20.52 ± 1.03% in submerged to helophyte *O. salzmannii*. This later species also presented the lowest levels of nitrogen and protein (0.46 ± 0.02 g% N and 2.85 ± 0.13 g% P, respectively) in marked contrast to *N. pulchella* that harbored the highest amounts (2.73 ± 0.03 g% N and 17.08 ± 0.16 g% P). Phenol levels also showed considerable amount variation, ranging from 3.84 ± 0.02 mg GAE/g extract in *X. brevifolia* to 77.98 ± 0.09 mg GAE/g extract in *T. domingensis*, representing helophyte and fixed emergent plants, respectively.
Table 2  Physical (% dry mass ash), nutritional (g% nitrogen, g% protein and % lipid), and general chemical defense (phenol content - mg GAE/g extract) properties of the studied aquatic plants. Values are means ± standard deviations. Different letters indicate statistically significant differences.

| Plant species          | Dry mass | Ash          | Lipid          | Nitrogen content | Protein content | Phenol content |
|------------------------|----------|--------------|----------------|------------------|-----------------|----------------|
| X. brevifolia          | 39.42 ± 0.47a | 16.76 ± 0.27a | –              | 0.94 ± 0.03a     | 5.85 ± 0.19a   | 3.84 ± 0.02a   |
| E. interstincta        | 28.24 ± 0.43b | 7.10 ± 0.18b  | 1.16 ± 0.07a   | 1.03 ± 0.04a     | 6.42 ± 0.25a   | 49.07 ± 0.06b  |
| T. domingensis         | 16.18 ± 0.21c | 8.33 ± 0.20c  | 6.65 ± 0.81bc  | 0.54 ± 0.04b     | 3.29 ± 0.22b   | 77.98 ± 0.09c  |
| P. illinoensis         | 7.50 ± 0.13d  | 14.88 ± 0.09d | 6.62 ± 0.20bc  | 2.00 ± 0.16c     | 12.48 ± 0.97c  | 63.99 ± 0.12d  |
| H. nymphoides          | 6.63 ± 0.05e  | 25.36 ± 0.21e | 7.85 ± 0.76bd  | 1.73 ± 0.10d     | 10.79 ± 0.63d  | –              |
| N. pulchella           | 9.42 ± 0.58f  | 9.36 ± 0.07f  | 1.44 ± 0.06e   | 1.62 ± 0.08e     | 10.15 ± 0.50fe | 70.85 ± 0.10c  |
| N. palustris           | 8.72 ± 0.18f  | 10.38 ± 0.17f | 1.84 ± 0.05f   | 2.73 ± 0.03f     | 17.08 ± 0.16f  | –              |
| E. azurea              | 9.32 ± 0.33h  | 22.97 ± 0.33h | 5.50 ± 0.50c   | 0.97 ± 0.02a     | 6.06 ± 0.13a   | 28.98 ± 0.02i  |
| S. auriculata          | 3.05 ± 0.09i  | 25.36 ± 0.17f | 9.48 ± 0.61d   | 1.62 ± 0.07d     | 10.13 ± 0.44d  | 41.77 ± 0.04f  |
| P. stratiotes          | –           | 39.35 ± 0.46f | 6.40 ± 0.41bc  | 2.46 ± 0.03f     | 15.35 ± 0.16f  | 60.11 ± 0.04b  |
| C. thalictroides       | –           | 15.87 ± 0.15j | 31.65 ± 1.52c  | 1.55 ± 0.03eh    | 9.67 ± 0.16c   | –              |
| O. salzmannii          | –           | 15.35 ± 0.09j | 20.52 ± 0.13f  | 0.46 ± 0.02b     | 2.85 ± 0.13b   | 55.72 ± 0.05i  |
| U. foliosa             | 3.49 ± 0.07j  | 21.39 ± 0.54l | 11.70 ± 0.70g  | 1.68 ± 0.11dh    | 10.52 ± 0.66dh | 47.04 ± 0.07j  |

Habit (life-form)

| SH    | 15.35 ± 0.09b | 20.52 ± 1.03a | 0.46 ± 0.02ab  | 2.85 ± 0.13ab    | 55.72 ± 0.05a  |
| H     | 39.42 ± 0.46a  | 16.76 ± 0.27a | –              | 0.94 ± 0.03ab    | 5.85 ± 0.19ab  | 3.84 ± 0.01b   |
| FE    | 16.18 ± 0.21b  | 8.33 ± 0.19b  | 6.56 ± 0.81ab  | 0.53 ± 0.03ab    | 3.29 ± 0.22b   | 77.98 ± 0.09b  |
| E     | 28.24 ± 0.42c  | 11.48 ± 4.80b | 16.40 ± 17.73a | 1.28 ± 0.29a     | 8.04 ± 1.79a   | 49.07 ± 0.06a  |
| EF    | 8.74 ± 0.85d   | 14.40 ± 5.60b | 3.85 ± 2.36bc  | 1.83 ± 0.67a     | 11.44 ± 4.18c  | 54.61 ± 19.45a |
| EFF   | 4.84 ± 1.97e   | 30.02 ± 7.00a | 7.91 ± 1.43bc  | 1.93 ± 0.40a     | 2.49 ± 0.83a   | 50.94 ± 10.04a |
| S     | 3.49 ± 0.07j   | 21.39 ± 0.54ab| 11.70 ± 0.70ab | 1.68 ± 0.10a     | 10.52 ± 0.38a  | 47.04 ± 0.07a  |

Relationships between plant traits and its consumption by B. glabrata

We did not uncover significant correlations between the susceptibility to consumption of fresh plants consumption of reconstituted plants, and the defensive capacity of the corresponding plant extracts (Pearson correlation coefficients, Fig. 3). Likewise, no significant relationships were observed between these three parameters and most of the various analyzed plant traits (ash, lipids, protein, total nitrogen and phenol) (Fig. 4), apart from a negative correlation between % extract inhibition and % lipids ($R_{Pearson} = -0.75; p = 0.02$). When we observe the PCA result, the first two axes of our PCA analysis summarized 76.78 % of the data variation ($PC1 = 44.19 \%$ and $PC2 = 32.59 \%$, Fig. 5A, Fig. 5B and Fig. 5C). The ordination of the samples (Fig. 5A and Fig. 5B) and variables (Fig. 5C) formed by these axes support our results described in previous sections. Overall, the PCA divided the samples into two groups: group 1 in the PC1 negative and PC2 positive quadrant; and group 2 in the PC1 positive and PC2 positive quadrant. In terms of species, the ordination of the of species sample (Fig. 5A), P. illinoensis, E. azurea and U. foliosa and formed group 1, and E. interstincta, X. brevifolia, T. domingensis, N. pulchella and H. nymphoides formed group 2. Ordination according to plant life-form (Fig. 5B) discriminated two groups, i.e., emergent floating and submerged plants (group...
1) and helophyte, fixed emergent and emergent plants (group 2). The ordination of the variables (loading, Fig. 5C) revealed that the group 1 encompassing plants whose extracts exhibited higher defensive activity against B. glabrata, exemplified by P. illionensis. The group 2 (Fig. 5C) represents the plants most susceptible to or preferred by B. glabrata as fresh or reconstituted food, in addition to revealing a strong correlation between dry mass and both consumption of fresh and reconstituted plants.

PCA analysis showed that the first two axes summarized 76.78% of the data variation (PC1 = 44.19% and PC2 = 32.59%, Figs. 5A, 5B, 5C). The ordination of the samples (Figs. 5A and 5B) and the variables in the space (Fig. 5C) formed by these axes reiterated the previous analysis of the data of the percentage of consumption of whole plants, reconstituted plants, inhibition of the extracts, and the several analyzed plant physical, nutritional and chemical traits and their correlations. Overall, the PCA divided the samples into two main groups: group 1 in the PC1 negative and PC2 in positive quadrant; and group 2 in the PC1 positive and PC2 positive quadrant. As main results, in the ordination of the species sample (Fig. 5A), it is possible to observe a group formed by P. illinoensis, E. azurea, U. foliosa and N. humboldtiana (group 1); and another group formed by E. interstincta, X. brevifolia, T. domingensis, N. pulchella and H. nymphoides (group 2). The ordination by life-form (Fig. 5B) discriminated two large plant groups: emergent floating and submerged plants (group 1) and helophyte, fixed emergent and emergent plants (group 2). The ordination of the variables (Fig. 5C) revealed that the group 1 includes those plants whose extracts exhibited higher defensive action against B. glabrata, mainly P. illinoensis. In contrast, group 2 (Fig. 5C) represents those plants most susceptible to or preferred by B. glabrata when offered to this gastropod either as a whole or as a reconstituted plant.

**Chemical profile of P. illinoensis extract**

The extract from P. illinoensis proved the most defensively capable against B. glabrata among the tested plants (Fig. 1C), so we further investigated its chemical profile by $^1$H...
Figure 4  Pearson correlation matrix pair about relationships between percentage consumption of whole plants, artificial foods, extract-mediated inhibition of consumption and the various physical, nutritional and chemical traits of the plant species. $r$ = Pearson correlation coefficient. In bold is highlighted the significant correlations.

NMR. Characteristic signals of fatty acids were observed in the aliphatic region, i.e., between $\delta$ 0.86 parts per million (ppm) and $\delta$ 1.70 ppm. An $\alpha$-CH$_2$ acyl group was observed at $\delta$ 2.32 (m), and strong overlapping resonance signals of methylene groups were also observed at $\delta$ 1.25 (sl). Additionally, we detected signals of exocyclic methylene ($\delta$ 4.0–5.0 ppm), methyl groups ($\delta$ 0.5–1.5 ppm) and hydrogen bonded to the sp$^2$ carbon of furane or lactone moieties ($\delta$ 6.09–7.11 ppm), which are characteristics of labdane diterpenes commonly reported in species of the *Potamogeton* genus (*Waridel et al.*, 2003) were also detected.

**DISCUSSION**

In this study, we investigated the susceptibility of various aquatic plants (as fresh and reconstituted artificial foods), to consumption by the generalist gastropod *B. glabrata*, and also explored the contribution of extracts as a chemical defense and several other plant traits to this plant-herbivore interaction.

We revealed rather indiscriminate consumption of all sampled plant species despite their varied life-forms, including emergent free floating (*H. nymphoides, S. auriculata*,
Figure 5  Results of Principal Component Analysis (PCA) separated by Dim 1 = PC1 and Dim 2 = PC2. (A) Two-dimensional score plot of the aquatic plant species. (B) Two-dimensional score plot of the different plant life-foms. (C) Two-dimensional loading plot. Black arrows represent the main quantitative variables (choice feeding assays) and gray arrows represent the supplementary quantitative variables (physical, nutritional and chemical traits).
P. stratiotes), emergent floating (E. azurea, N. pulchella, P. illinoensis), emergent (E. interstincta and C. thalictroides), fixed emergent (T. domingensis), submerged to helophyte (O. salzmannii), and helophyte (X. brevifolia). This outcome contrasts with several previous studies that have reported how the life-form of aquatic plants impacts their relationships with consumers (Lodge, 1991; Smolders et al., 2000; Bakker et al., 2016). Submerged, emergent, and floating-leaved plants are subjected to high herbivory pressure by diverse consumers (Lodge, 1991), implying that different aquatic plant life-forms predispose to different consumer interactions. For example, both emergent and shallow submerged aquatic plants would be accessible to terrestrial herbivores, whereas obligate aquatic herbivores could only consume submerged plants (Gross & Baker, 2012). Previously, herbivory rates were found to be lower on emergent aquatic plants relative to submerged ones (Bakker et al., 2016), and emergent plants appear to be less susceptible to consumption by invertebrates than submerged and floating forms (Lodge, 1991). An analysis of phenolic contents in 40 aquatic and semi-aquatic plants revealed lower levels in submerged species compared to emergent or floating ones (Smolders et al., 2000), although that pattern was not observed for the 10 species we analyzed here.

The dry mass contents of aquatic plants have been negatively correlated with consumption rates (e.g., Elger & Willby, 2003; Wong et al., 2010) and structural defenses are more commonly found among upland plants than wetland ones (Lodge, 1991), though either of these features explain preferential consumption of aquatic plants (Prejs, 1984). Among the plants with higher amounts of dry matter considered in our study, E. interstincta, X. brevifolia and T. domingensis, only in the emergent T. domingensis might this trait play any importance in its consumption, albeit perhaps not statistically significantly. Typha domingensis was largely rejected as food by B. glabrata when presented as fresh plant, but it was the most consumed when presented as reconstituted artificial food. Interestingly, the three emergent species most consumed by B. glabrata, E. interstincta > N. pulchella > H. nymphoides, exhibited low levels of ash and high amounts of aerenchyma (T. Konno, 2021, pers. comm.), with this latter being an important characteristic facilitating the plant fluctuation, but potentially rendering them more susceptible to herbivory.

In general, submerged plants need fewer structural components, but generate higher nitrogen levels and suffer higher rates of herbivory relative to emergent species (Bakker et al., 2016). In our experiments using reconstituted plants, the submerged species O. salzmannii that displays low nutritional value (low amounts of nitrogen and protein) was moderately consumed by B. glabrata, but its extract stimulated the consumption by the gastropod. In contrast, U. foliosa has high levels of nitrogen but was largely not consumed by B. glabrata.

We found no evidence of a relationship between the susceptibility to consumption of fresh plants and their reconstituted artificial foods or for the defensive action of their extracts on this gastropod. These results support consumption of the assayed plants by B. glabrata being rather generalized, as they have been reported for a variety of other aquatic consumers, such as gastropods (Sheldon, 1987), crayfishes (Seroll & Coler, 1975; Cronin et al., 2002), grass carp (Wiley, Pescitelli & Wike, 1986; Bowers, Pauley & Thomas, 1987; Parker & Hay, 2005), and mosse (Fraser, Chavez & Paloheimo, 1984).
The evidence clearly points to *P. illinoensis* being the least susceptible to consumption, either as fresh plant or as reconstituted food, and its extract showed the highest percentage of inhibition among all the species we studied. This outcome supports several previous studies showing that aquatic plants are commonly unpalatable and contain a variety of secondary metabolites (e.g., terpenes, phenols and alkaloids) known for their ecological role as chemical deterrents against herbivory (*Ostrofsky & Zettler, 1986; Kubanek et al., 2001; Cronin et al., 2002; Prusak, O’Neal & Kubanek, 2005). *Potamogeton illinoensis* is a known producer of secondary metabolites (*Zhang et al., 2011*) and its close relative *Potamogeton pectinatus* exhibit algicidal and allelopathic activity (*Waridel et al., 2003*). Consequently, these chemicals likely exhibit ecological roles as deterrent to the activity of herbivores such as the gastropod *B. glabrata*.

Several other species we studied also presented high percentage of inhibition of *B. glabrata* consumption, such as *X. brevifolia, T. domingensis, E. aurea, N. humboldtiiana, U. foliarea, E. interstincta* and *P. stratoideas*. Several of these or related species are known to produce secondary metabolites. For example, *T. domingensis* is known to produce phenols and fatty acids (*Gallardo-Williams et al., 2002*), and an extract from it inhibits feeding by crayfish (*Bolser et al., 1998*). Species of *Xyris* also produce defensive chemicals, such as the isocoumarins of *X. indica* (*Ruangrungsi et al., 1995*), the anthraquinones of *X. semifuscata* (*Fournier et al., 1975*) and the flavonoids from *X. itatiayensis, X. longiscapa* and *X. obtusiuscula* (*Varanda, Rondinoni & dos Santos, 2002*). Phenols, many displaying allelopathic activity, have been identified in species of *Eleocharis* (*Frank & Dechoretz, 1980; Zaman et al., 2018*), and *Eichornia crassipes* produces phenalenone compounds and sterols (*Lalitha, Sripathi & Jayanthi, 2012*). Anthraquinones (*Siva et al., 2012*) and cyclic peptides (*Seydel & Dörnenburg, 2006*) have been identified from the genus *Oldenlandia*. Similarly, several classes of secondary metabolites have been detected in *Ceratopteris thalictroides* (*Smitha & Vadivel, 2019*), and autoinhibitory compounds were also identified in *Salvinia molesta* (*Li et al., 2016*). Chemicals from *Potamogeton maackianus* were shown to exert allelopathic activity against the microalga *Selenastrum capricornutum* (*Zhang et al., 2011*), and *ent*-labdane diterpenes from *P. pectinatus* had the same effect on *Raphidocelis subcapitata* (*Waridel et al., 2003*). Interestingly, we identified en-labdane diterpenes in the *P. illinoensis* extract we assessed by 1H NMR. Alkaloids have been detected in *P. amplifolius* (*Ostrofsky & Zettler, 1986*), and high levels of phenols were found in the propagules of three species of *Potamogeton* (*P. gramineus, P. nodosus* and *P. pectinatus*) (*Spencer & Ksander, 1994*), proving evidence that species of this genus protect themselves from herbivory through chemical defense. Based on our findings, we suggest that, apart from *H. nymphoides*, all other assayed species are producers of chemicals that could act as chemical deterrents, resulting in those species being less susceptible to consumption by *B. glabrata*.

The defensive function against herbivory of phenols in plants is well known (*Lodge, 1991*). Almost all of the species we assayed produce phenols, such as *E. interstincta, N. humboldtiiana, O. salzmannii, P. stratoites* and *T. domingensis*, but these plants were well-consumed by *B. glabrata*. The only exception was *P. illinoensis*, representing the plant least consumed by *B. glabrata*, but we found it harbored other defensive substances, such as diterpenoids.
Plant nutritional value and chemical defense capability are the primary traits determining plant palatability (Cronin et al., 2002; Sotka et al., 2009). Aquatic plants of high nutritional value should be favored by generalist herbivores such as crayfish and apple snails (Ampullariidae) because eating them would maximize fitness, yet there is a trade-off between feeding on nutrient-rich food and ingesting defensive plant metabolites (Lodge, 1991). However, two plant species with the highest nutritional values (nitrogen and protein) that we studied, *N. pulchella* and *P. stratiotes*, were not the most preferred by *B. gabrata* in our feeding assays. Floating plants, typically exhibit higher nitrogen contents than other aquatic plants (Lodge, 1991), but in four of the floating plants we analysed (*H. nymphoides*, *N. humboldtiana*, *N. pulchella* and *P. stratiotes*), intermediate-to-high nitrogen and protein levels were detected, but they were not preferred food items of *B. glabrata*.

Despite all the plant parameters assayed in this study, we did not detect prevalence of one protective or attractive factors that could explain the plant-herbivore relationship with *B. glabrata*, despite all of the traits we analyzed being known to influence feeding by invertebrates on aquatic plants. This outcome may not be surprising, since the consumption of aquatic plants may be unrelated to their physical and chemical characteristics (Lodge, 1991). However, we believe that further studies on the relationships of aquatic plants with herbivores are still needed, necessitating an integrative approach as applied herein and assessing other types of herbivores. Although gastropods exert an important top-down control on aquatic plants through grazing (Lodge, 1991; Newman, 1991; Lodge et al., 1998), this effect is dispersed across the food web due to their generalist diet. Despite identifying a hierarchy of plant consumption by *B. glabrata*, its lack of a clear preference implies that this gastropod makes its food choice based on multiple plant traits, thereby exerting a diffuse effect on these plants in their natural environment.

**ADDITIONAL INFORMATION AND DECLARATIONS**

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**Competing Interests**
The authors declare there are no competing interests.

**Author Contributions**
- Renato Crespo Pereira and Tatiana U.P. Konno analyzed the data, authored or reviewed drafts of the paper, and approved the final draft.
- Nathália Nocchi conceived and designed the experiments, performed the experiments, analyzed the data, prepared figures and/or tables, authored or reviewed drafts of the paper, and approved the final draft.
- Angelica R. Soares conceived and designed the experiments, performed the experiments, analyzed the data, authored or reviewed drafts of the paper, and approved the final draft.

**Ethics**
The following information was supplied relating to ethical approvals (i.e., approving body and any reference numbers):
- Sistema de Autorização e Informação em Biodiversidade - SISBio, Instituto Chico Mendes de Preservação da Biodiversidade approved this study (SISBio 30299).

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**Data Availability**
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**REFERENCES**

Bakker ES, Wood KA, Pagès JF, Veend (Ciska) GF, Christianene MJA, Santamaríaf L, Nolet BA, Hilt S. 2016. Herbivory on freshwater and marine macrophytes: a review and perspective. *Aquatic Botany* **135**:18–36 DOI 10.1016/j.aquabot.2016.04.008.

Barbarino E, Lourenço SO. 2005. An evaluation of methods for extraction and quantification of protein from marine macro- and microalgae. *Journal of Applied Phycology* **17**:447–460 DOI 10.1007/s10811-005-1641-4.

Bolser RC, Hay ME. 1998. A field test of inducible resistance to specialist and generalist herbivores using the water lily *Nuphar luteum*. *Oecologia* **116**:143–153 DOI 10.1007/s004420050573.
Bolser RC, Hay ME, Lindquist N, Fenical W, Wilson D. 1998. Chemical defenses of freshwater macrophytes against crayfish herbivory. *Journal of Chemical Ecology* 24:1639–1658 DOI 10.1023/A:1020816511924.

Bowers KL, Pauley GB, Thomas GL. 1987. Feeding preference on Pacific Northwest aquatic plant species by diploid and triploid grass carp (*Ctenopharyngodon idella*). In: *Proceedings of 21st annual meeting, aquatic plant control research group. Misc. Pap. A-87-2*. Vicksburg, MS: US Army Engineer Waterways Experiment Station, 133–140.

Chambers PA, Hanson JM, Prepas EE. 1991. The effect of aquatic plant chemistry and morphology on feeding selectivity by the crayfish, Orconectes virilis. *Freshwater Biology* 25:339–348 DOI 10.1111/j.1365-2427.1991.tb00496.x.

Conklin-Brittain NL, Dierenfeld ES, Wrangham RW, Norconk M, Silver SC. 1999. Chemical protein analysis: a comparison of Kjeldahl crude protein and total ninhydrin protein from wild, tropical vegetation. *Journal of Chemical Ecology* 25:2601–2622 DOI 10.1023/A:1020835120701.

Cook CDK. 1996. *Aquatic plant book*. Amsterdan: SPB Academic Publishing.

Cronin G. 1998. Influence of macrophyte structure, nutritive value, and chemistry on the feeding choices of a generalist crayfish. In: Jeppesen E, Søndergaard M, Søndergaard M, Christoffersen K, eds. *The structuring role of submerged macrophytes in lakes*. New York: Springer-Verlag, 307–317.

Cronin G, Hay ME. 1996. Amphipod grazing induces chemical defenses in the brown alga *Dictyota menstrualis*. *Ecology* 77:2287–2301 DOI 10.2307/2265731.

Cronin G, Lodge DM, Hay ME, Miller M, Hill AM, Horvath T, Bolser RB, Lindquist N, Wahl MT. 2002. Crayfish feeding preferences for freshwater macrophytes: the influence of plant structure and chemistry. *Journal of Crustacean Biology* 22:708–718 DOI 10.1163/20021975-99990285.

Dorgelo J, Heykoop M. 1985. Avoidance of macrophytes by *Daphnia longispina*. *Verh. Internationale Vereinigung für Theoretische Und Angewandte Limnologie: Verhandlungen*.

Doucet CM, Fryxell JM. 1993. The effect of nutritional quality on forage preference by beavers. *Oikos* 67:201–208 DOI 10.2307/3545464.

Elger A, Lemoine D. 2004. Determinants of macrophyte palatability to the pond snail *Lymnaea stagnalis*. *Freshwater Biology* 50(1):86–95 DOI 10.1111/j.1365-2427.2004.01308.x.

Fornoff F, Gross EM. 2014. Induced defense mechanisms in an aquatic angiosperm to insect herbivory. *Oecologia* 175:173–185 DOI 10.1007/s00442-013-2880-8.

Fournier G, Bercht CAL, Paris RR, Paris MR. 1975. 3- Methoxychrysazin, a new antraquinone from *Xyris semifusca*. *Phytochemistry* 14:2099 DOI 10.1016/0031-9422(75)83146-3.

Frank PA, Dechoretz N. 1980. Allelopathy in dwarf spikerush (*Eleocharis coloradoensis*). *Weed Science* 28:499–505 DOI 10.1017/S0043174500061105.

Fraser D, Chavez ER, Paloheimo JE. 1984. Aquatic feeding by moose: selection of plant species and feeding areas in relation to chemical composition and characteristics of lakes. *Canadian Journal of Zoology* 62:80–87 DOI 10.1139/z84-014.
Gallardo-Williams MT, Geiger CL, Pidala JA, Martin DF. 2002. Essential fatty acids and phenolic acids from extracts and leachates of southern cattail (Typha domingensis P.). *Phytochemistry* 59:305–308 DOI 10.1016/s0031-9422(01)00449-6.

Gross EM, Baker ES. 2012. The role of plant secondary metabolites in freshwater macrophyte–herbivore interactions: limited or unexplored chemical defences? In: Iason GR, Dockie M, Hartley SE, eds. *The ecology of plant secondary metabolites: from genes to global processes. British ecological society*, Cambridge: Cambridge University Press, 154–169.

Gutierrez MF, Paggi JC. 2014. Chemical repellency and adverse effects of free-floating macrophytes on the cladoceran Ceriodaphnia dubia under two temperature levels. *Limnology* 15:37–45 DOI 10.1007/s10201-013-0411-2.

Hairston NG, Smith FE, Slobodkin LB. 1960. Community structure, population control, and competition. *American Naturalist* 94:421–425 DOI 10.1086/282146.

Hay ME, Kappel QE, Fenical W. 1994. Synergisms in plant defenses against herbivores: interactions of chemistry, calcification and plant quality. *Ecology* 75:1714–1726 DOI 10.2307/1939631.

Jamovi The Project. 2021. Jamovi. (Version 1.6). Available at https://www.jamovi.org.

Jones JB. 2001. *Laboratory guide for conducting soil tests and plant analysis*. Boca Raton: CRC Press.

Kreider JL, Watts SA. 1998. Behavioral (feeding) responses of the crayfish, Procambarus clarkii, to natural dietary items and common components of formulated crustacean feeds. *Journal of Chemical Ecology* 24:91–111 DOI 10.1023/A:1022341013148.

Kubanek J, Hay ME, Brown PJ, Lindquist N, Fenical W. 2001. Lignoid chemical defenses in the freshwater macrophyte Saururus cernuus. *Chemoecology* 11:1–8 DOI 10.1007/PL00001826.

Lalitha P, Sripathi SK, Jayanthi P. 2012. Secondary metabolites of Eichornia crassipes (Waterhyacinth): a review (1949 to 2011). *Natural Products Communication* 7:1249–1256 DOI 10.1177/1934578X1200700939.

Li S, Wang P, Yuan W, Su Z, Bullard SH. 2016. Endocidal regulation of secondary metabolites in the producing organisms. *Scientific Reports* 6:29315 DOI 10.1038/srep29315.

Lockwood III JR. 1998. On the statistical analysis of multiple-choice feeding preference experiments. *Oecologia* 116:475–481 DOI 10.1007/s004420050612.

Lodge DM. 1991. Herbivory on freshwater macrophytes. *Aquatic Botany* 41:195–224 DOI 10.1016/0304-3770(91)90044-6.

Lodge DM, Cronin G, Van Donk E, Froelich AJ. 1998. Impact of herbivory on plant standing crop: comparison among biomes, between vascular and nonvascular plants, and among freshwater herbivore taxa. In: Jeppesen E, Sondergaard M, Sondergaard M, Christoffersen K, eds. *The structuring role of submerged macrophytes in lakes. Ecological studies*, vol. 131. New York: Springer, 149–174.

Lubchenco J. 1978. Plant species diversity in a marine intertidal community: importance of herbivore food preference and algal competitive abilities. *American Naturalist* 112:23–39 DOI 10.1086/283250.
Morrison WE, Hay ME. 2011. Induced chemical defenses in a freshwater macrophyte suppress herbivore fitness and the growth of associated microbes. *Oecologia* 165:427–436 DOI 10.1007/s00442-010-1791-1.

Newman RM. 1991. Herbivory and detritivory on fresh-water macrophytes by invertebrates: a review. *Journal of the North American Benthological Society* 10:89–114 DOI 10.2307/1467571.

Ostrofsky ML, Zettler ER. 1986. Chemical defences in aquatic plants. *Journal of Ecology* 74:279–287 DOI 10.2307/2260363.

Parker JD, Collins DO, Kubanek J, Sullards MC, Bostwick D, Hay ME. 2006. Chemical defenses promote persistence of the aquatic plant *Micranthemum umbrosum*. *Journal of Chemical Ecology* 32:815–833 DOI 10.1007/s10886-006-9038-7.

Parker J, Hay ME. 2005. Biotic resistance to plant invasions? Native herbivores prefer non-native plants. *Ecology Letters* 8:959–967 DOI 10.1111/J.1461-0248.2005.00799.X.

Pennak RW. 1973. Some evidence for aquatic macrophytes as repellents for a limnetic species of *Daphnia*. *Internationale Revue der Gesamten Hydrobiologie und Hydrographie* 58:569–576 DOI 10.1002/iroh.19730580406.

Prejs A. 1984. Herbivory by temperate freshwater fishes and its consequences. *Environmental Biology of Fishes* 10:281–296 DOI 10.1007/BF00001481.

Prusak AC, O’Neal J, Kubanek J. 2005. Prevalence of chemical defenses among freshwater macrophytes. *Journal of Chemical Ecology* 31:1145–1160 DOI 10.1007/s10886-005-4253-1.

R Core Team. 2020. R: a Language and environment for statistical computing. (Version 4.0). (R packages retrieved from MRAN snapshot 2020-08-24). Available at https://cran.r-project.org.

Ramesh S, Rajan R, Santhanan R. 2014. *Freshwater phytopharmaceutical compounds*. Boca Raton: CRC Press.

Ramluckana K, Moodley KG, Bux F. 2014. An evaluation of the efficacy of using selected solvents for the extraction of lipids from algal biomass by the soxhlet extraction method. *Fuel* 116:103–108 DOI 10.1016/j.fuel.2013.07.118.

Rothhaupt KO, Fornoff F, Yohannes E. 2015. Induced responses to grazing by an insect herbivore (*Acentria ephemerella*) in an immature macrophyte (*Myriophyllum spicatum*): an isotopic study. *Ecology & Evolution* 5:3657–3665 DOI 10.1002/ece3.1624.

Ruangrunsgi N, Toshikazu S, Phadungcharoent T, Suriyagan S, Murakoshi I. 1995. Isocoumarins from *Xyris indica*. *Phytochemistry* 38:481–483 DOI 10.1016/0031-9422(94)00714-5.

Seroll A, Coler RA. 1975. Demonstrated food preferences of *Orconectes immunis* (Hagen) (Decapoda, Astacidea). *Crustaceana* 29:319–320 DOI 10.1163/156854075X00469.

Seydel P, Dörnenburg H. 2006. Establishment of in vitro plants, cell and tissue cultures from *Oldenlandia affinis* for the production of cyclic peptides. *Plant Cell, Tissue Organ Culture* 85:247–255 DOI 10.1007/s11240-005-9056-0.

Sheldon SP. 1987. The effects of herbivorous snails on submerged macrophyte communities in Minnesota lakes. *Ecology* 68:1920–1931 DOI 10.2307/1939883.
Siva R, Mayes S, Behera SK, Rajasekarana C. 2012. Anthraquinones dye production using root cultures of Oldenlandia umbellata L. *Industrial Crops and Production* **57**:415–419 DOI 10.1016/j.indcrop.2011.12.027.

Smitha V, Vadivel V. 2019. Phytochemical screening for active compounds in Ceratopteris thalictroides (L.) Brogn. *Journal of Pharmacognosy and Phytochemistry* **8**:3556–3559.

Smolders AJP, Vergeer LHT, Van der Velde G, Roelofs JGM. 2000. Phenolic contents of submerged, emergent and floating leaves of aquatic and semi-aquatic macrophyte species: why do they differ? *Oikos* **91**:307–310 DOI 10.1034/j.1600-0706.2000.910211.x.

Sotka EE, Forbey J, Horn M, Poore AG, Raubenheimer D, Whalen KE. 2009. The emerging role of pharmacology in understanding consumer-prey interactions in marine and freshwater systems. *Integrative and Comparative Biology* **49**:291–313 DOI 10.1093/icb/icp049.

Spencer DF, Ksander GG. 1994. Phenolic acid content of vegetative propagules of *Potamogeton* spp. and *Hydrilla verticillata*. *Journal of Aquatic Plant Management* **32**:71–73.

Thomas JD, Nwanko DI, Sterry PR. 1985. The feeding strategies of juvenile and adult *B. glabrata* (Say) under simulated natural conditions and their relevance to ecological theory and snail control. *Proceedings of the Royal Society of London. Series B: Biological Sciences* **226**:117–209 DOI 10.1098/rspb.1985.0090.

Vaissie P, Monge A, Husson F. 2021. Factoshiny: perform factorial analysis from FactoMineR with a shiny application. R package version 2.4. Available at https://cran.r-project.org/web/packages/Factoshiny/Factoshiny.pdf.

Varanda EM, Rondinoni C, dos Santos DYAC. 2002. Flavonoids from *Xyris* species (Xyridaceae). *Biochemical Systematics and Ecology* **30**:997–998 DOI 10.1016/S0305-1978(02)00035-2.

Vermeij GJ. 2016. Plant defences on land and in water: why are they so different? *Annals of Botany* **117**:1099–1109 DOI 10.1093/aob/mcw061.

Waridel P, Wolfender J-L, Lachavanne J-B, Hostettmann K. 2003. ent-Labdane diterpenes from the aquatic plant *Potamogeton pectinatus*. *Phytochemistry* **64**:1309–1317 DOI 10.1016/j.phytochem.2003.08.014.

Wiley MJ, Pescitelli SM, Wike LD. 1986. The relationship between feeding preferences and consumption rates in grass carp and grass carp × bighead carp hybrids. *Journal of Fish Biology* **29**:507–514 DOI 10.1111/j.1095-8649.1986.tb04966.x.

Wong PK, Liang Y, Liu NY, Qiu J-W. 2010. Palatability of macrophytes to the invasive freshwater snail *Pomacea canaliculata*: differential effects of multiple plant traits. *Freshwater Biology* **55**:2023–2031 DOI 10.1111/j.1365-2427.2010.02458.x.

Zaman F, Iwasaki A, Suenaga K, Kato-Noguchi K. 2018. Allelopathic property and an allelopathic substance in *Eleocharis atropurpurea* (Retz.). *Theoretical and Experimental Plant Physiology* **30**:347–355 DOI 10.1007/s40626-018-0130-5.
Zhang SH, Cheng SP, Sun PS, Wwng HQ, Wu ZB. 2011. Isolation and identification of antialgal compounds from *Potamogeton maackianus* by activity-guided fractionation. *Allelopathy Journal* 28:95–104.

Zhang M, Guo B, Zhang R, Chi J, We Z, Xu Z, Zhang Y, Tang X. 2006. Separation, purification and identification of antioxidant compositions in black rice. *Agricultural Sciences in China* 5:431–440 DOI 10.1016/S1671-2927(06)60073-4.