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Authors
Chevrier, Jonathan
Gunier, Robert B
Bradman, Asa
et al.

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Maternal Urinary Bisphenol A during Pregnancy and Maternal and Neonatal Thyroid Function in the CHAMACOS Study

Jonathan Chevrier,¹ Robert B. Gunier,¹ Asa Bradman,¹ Nina T. Holland,¹ Antonia M. Calafat,² Brenda Eskenazi,¹ and Kim G. Harley¹

¹Center for Children’s Environmental Health Research, School of Public Health, University of California, Berkeley, Berkeley, California, USA; ²Division for Laboratory Sciences, National Center for Environmental Health, Centers for Disease Control and Prevention, Atlanta, Georgia, USA

BACKGROUND: Bisphenol A (BPA) is widely used in the manufacture of polycarbonate plastic bottles, food and beverage can linings, thermal receipts, and dental sealants. Animal and human studies suggest that BPA may disrupt thyroid function. Although thyroid hormones play a determinant role in human growth and brain development, no studies have investigated relations between BPA exposure and thyroid function in pregnant women or neonates.

OBJECTIVE: Our goal was to evaluate whether exposure to BPA during pregnancy is related to thyroid hormone levels in pregnant women and neonates.

METHODS: We measured BPA concentration in urine samples collected during the first and second half of pregnancy in 476 women participating in the CHAMACOS (Center for the Health Assessment of Mothers and Children of Salinas) study. We also measured free thyroxine (T₄), total T₄, and thyroid-stimulating hormone (TSH) in women during pregnancy, and TSH in neonates.

RESULTS: Associations between the average of the two BPA measurements and maternal thyroid hormone levels were not statistically significant. Of the two BPA measurements, only the one taken closest in time to the TH measurement was significantly associated with a reduction in total T₄ (β = –0.13 μg/dL per log₂ unit; 95% CI: –0.25, 0.00). The average of the maternal BPA concentrations was associated with reduced TSH in boys (β = 9.9% per log₂ unit; 95% CI: –15.9%, –3.5%) but not in girls. Among boys, the relation was stronger when BPA was measured in the third trimester of pregnancy and decreased with time between BPA and TH measurements.

CONCLUSION: Results suggest that exposure to BPA during pregnancy is related to reduced total T₄ in pregnant women and decreased TSH in male neonates. Findings may have implications for fetal and neonatal development.

KEY WORDS: bisphenol A, endocrine disruption, neonates, pregnancy, thyroid hormone.

Bisphenol A (BPA) is widely used in the manufacture of polycarbonate plastic bottles, epoxy resins used in the inner lining of food and beverage cans, thermal receipts, medical equipment, tableware, and water supply pipes. Approximately 2.4 billion pounds of BPA were produced in the United States in 2007 [U.S. Environmental Protection Agency (EPA) 2010], and 95% of U.S. women of reproductive age (18–44 years) from the 2007–2008 National Health and Nutrition Examination Survey (NHANES) had detectable BPA levels in their urine [Centers for Disease Control and Prevention (CDC) 2011]. Unconjugated BPA has been detected in cord blood, placental tissue, and amniotic fluid, suggesting that the chemical can cross the placenta (Ikekuzi et al. 2002; Schonfelder et al. 2002). Following a review of BPA’s potential to cause adverse reproductive and developmental effects, the National Toxicology Program (NTP) published a report in 2008 expressing “some concern” (the midpoint of a five-level scale) that current human exposure to BPA resulted in adverse effects on brain development and behavior in fetuses, infants, and children (Chapin et al. 2008).

Thyroid hormones (TH) play an essential role in pre- and postnatal growth and brain development in humans. Although severe maternal and neonatal thyroid insufficiency or overactivity has been known to alter cognition, behavior, and growth for more than a century (Dunn 1993), more recent evidence suggests that mild alterations in thyroid function may also influence these outcomes (Pop et al. 1999). Potential effects of BPA on cognition and behavior may thus be attributable in part to disruption of thyroid function.

Experimental evidence offers some support for this hypothesis. For instance, one in vitro study reported that BPA antagonized the ability of TH to affect oligodendrocyte differentiation (Seiwa et al. 2004). Moriyma et al. (2002) also found that BPA binds to the thyroid receptor (TR), antagonizes triiodothyronine (T₃) binding to the TR, and inhibits TR-mediated gene expression in vivo. Furthermore, an animal study suggested that oral exposure to BPA results in a nonmonotonic transitory decrease in free, but not total, thyroxine (T₄) in pregnant rats (Xu et al. 2007). Prenatal exposure to BPA, on the other hand, was related to a transitory dose-related elevation in total T₄ among both male and female pups in one study (Zoeller et al. 2005) and in a nonmonotonic increase in free T₄ (at postnatal day 7) followed by a decrease (at postnatal day 21) among male pups only in another study (Xu et al. 2007); a third study found no effect of prenatal BPA exposure on total T₄ (Kobayashi et al. 2005). The studies by Zoeller et al. (2005) and Xu et al. (2007) found effects at the lowest doses administered (1 mg/kg body weight and 0.1 mg/L drinking water, respectively). Studies conducted in nonpregnant animals generally found no effect of BPA on thyroid hormone levels (Nieminen et al. 2002; Seidlova-Wuttke et al. 2005).

Four human studies have examined relations between exposure to BPA and thyroid function, yielding conflicting results. Meeker et al. (2010) found no association between serum free T₄, total T₃, and thyroid-stimulating hormone (TSH) and BPA concentrations in urine samples collected the same day in 167 men seeking treatment at an infertility clinic in Boston, Massachusetts. However, when BPA urinary concentrations in samples collected at multiple time points (1–3 measurements taken 3–75 days apart) were considered, their geometric means (GM) were associated with depressed TSH. In addition, BPA urinary concentrations were positively associated with free T₃ among factory workers with high exposure.

Address correspondence to J. Chevrier, Center for Environmental Research and Children’s Health (CERCH), School of Public Health, University of California, Berkeley, 1995 University Ave., Suite 265, Berkeley, CA 94704-7392 USA. Telephone: (510) 642-8917. Fax: (510) 642-9083. E-mail: chevrier@berkeley.edu

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We collected data as part of the Center for the Health Assessment of Mothers and Children of Salinas (CHAMACOS), a longitudinal birth cohort study of environmental exposures and health among pregnant women and children. Women were eligible to participate in this study if they spoke Spanish or English, were ≥ 18 years old, had completed < 20 weeks gestation, and qualified for California’s low-income health insurance program. We enrolled participants (n = 601) between October 1999 and October 2000. We excluded women from the present analyses if they had twins (n = 5), had a miscarriage (n = 20) or stillborn baby (n = 3), were lost to follow-up (n = 40), were taking medication that could affect TH levels (n = 1), or did not have a BPA urine measurement during pregnancy (n = 28). TH was not measured in 169 mothers due to insufficient serum volume, leaving a final sample of 335 for the maternal analyses. For the analyses of neonatal TSH, we excluded 2 neonatal deaths and 138 neonates for whom TSH measures (n = 114) or age at heel stick (n = 24) were missing from medical records. A total of 364 mother–child pairs were included in the neonatal analyses; 476 women were included in at least one analysis. We obtained written informed consent from all participants and all research was approved by the University of California (UC), Berkeley, Committee for the Protection of Human Subjects before commencement of the study. CDC relied on the UC Berkeley Committee.

**Interviews.** We interviewed participants at enrollment (~ 13 weeks gestation) and near the end of the second trimester (~ 26 weeks gestation) using structured questionnaires. We collected information on demographics including maternal age, race/ethnicity, education, country of birth, and time lived in the United States. We also collected information on smoking, alcohol consumption, and drug use during pregnancy. In addition, we obtained data on iodine intake during pregnancy using a modified version of the Spanish-language Block 98 Questionnaire (Block et al. 1990; Harley et al. 2005).

**Sample analysis.** Measurement of thyroid hormone. We collected maternal blood samples during the second interview and processed them immediately after collection. Samples were stored at –80°C at the UC Berkeley School of Public Health Biorepository until analysis. A pilot study showed that the number of freeze–thaw cycles was related to an increase in free T\textsubscript{4} (data not shown). Samples were thus thawed only once, shipped refrigerated and analyzed for TH by Quest Diagnostics’ Nichols Institute (San Juan Capistrano, CA) within 48 hr.

Free T\textsubscript{4} was measured using direct equilibrium dialysis followed by radioimmunoassay (Nelson and Tomei 1988), which provides accurate measurements despite pregnancy-induced elevations in T\textsubscript{4}-bound proteins (Nelson et al. 1994). Total T\textsubscript{4} and TSH were measured in maternal serum using solid-phase immunochemiluminometric assays (Bayer ADVIA Centaur system; Siemens Healthcare Diagnostics, Deerfield, IL). The limits of detection (LODs) were 0.1 ng/dL for free T\textsubscript{4}, 0.1 µg/L for total T\textsubscript{4}, and 0.01 mIU/L for TSH.

Neonatal TSH is routinely measured by the California Department of Public Health by graphite furnace atomic absorption spectrophotometry. Serum polychlorinated biphenyls (PCBs), hexachlorobenzene (HCB), and polybrominated diethyl ether (PBDE) flame retardants were measured by the CDC using gas chromatography/isotope-dilution high-resolution mass spectrometry (Castorina et al. 2011; Sjödin et al. 2004). We collected samples for these analyses at the end of the second trimester of gestation (mean ± SD = 27 ± 3 weeks gestation).

**Statistical analysis.** We used analysis of variance (ANOVA) to compare BPA urinary concentrations across variable categories and Pearson’s correlations to evaluate bivariate associations between continuous variables. Because animal studies have suggested non-mono tonic dose responses between BPA and TH levels (Xu et al. 2007), we ran generalized additive models (GAM) with a 3-degrees-of-freedom cubic spline including covariates selected for the final models, to evaluate the shape of exposure–response curves in our study population. None of the tests for digression from linearity were significant (p < 0.15), suggesting that relations did not depart from linearity. We therefore included linear terms for BPA in multiple linear regression models. We expressed exposure as the average of the two (first and second half of pregnancy) BPA measurements. BPA urinary concentrations were heavily right-skewed and thus were log\textsubscript{10}-transformed to reduce the influence of outliers. We log\textsubscript{10}-transformed TSH to normalize residuals; free and total T\textsubscript{4} were expressed on the arithmetic scale. Regression coefficients were adjusted for the following variables:

- gestational age (continuous, linear and quadratic terms)
- maternal age, parity, income, race, and education
- smoking, alcohol use, and caffeine use during pregnancy
- maternal TSH at the second trimester of pregnancy
- pre-pregnancy BMI
- maternal body weight at the second trimester of pregnancy
- alcohol use during pregnancy
- fruit, vegetable, and fish consumption during pregnancy
- ethnic group
- living environment
- maternal thyroid function
- other environmental exposures (e.g., metals, organochlorines, and pesticides)

These variables were selected based on previous research and were included in models to account for potential confounding factors.
thus represent mean (free and total T₄) or percent (TSH) change in outcomes for each doubling in BPA concentration. In addition, we ran multiple logistic regressions, categorizing TH as normal versus above or below their respective reference ranges.

We considered the following variables that may influence TH levels as potential confounders (expressed as shown in Table 1 or in parentheses): maternal age (continuous), race/ethnicity, education, family income, country of birth, number of years spent in the United States, parity, gestational age at the time of blood collection (for maternal TH analyses only; weeks, continuous), iodine intake (continuous), and smoking, alcohol consumption, and illegal drug use during pregnancy. For neonatal TSH analyses, we also considered sex, delivery mode, and age at the time of heel stick (hours, continuous). TSH surges at birth and sharply declines during the first few days of life (Brown et al. 2005). Our previous work suggests that age at the time of TSH measurement is a strong confounder for associations with other environmental contaminants in the CHAMACOS population (Chevrier et al. 2007, 2011). We thus used GAMs and applied cubic splines to this variable to minimize residual confounding. Finally, based on results from prior studies in this population (Chevrier et al. 2007, 2008, 2010), we considered environmental exposures such as (log₁₀-transformed) blood lead (micrograms per deciliter), total serum PCBs, HCB, and PBDE flame retardants (nanograms per gram lipids) for inclusion in models. We included covariates in final models if they were associated with both TH and BPA (p < 0.20) in bivariate analyses. For analyses of maternal TH, final models comprised mother’s age, education level, country of birth, poverty level, alcohol and drug use during pregnancy, iodine intake, and HCB and PCB serum concentrations. Final models of neonatal TSH included mother’s country of birth and child’s age at thyroid hormone measurement.

BPA has a short half-life in humans (< 6 hr) (Volkel et al. 2002), and some experimental studies suggest that exposure to BPA may have a transitory effect on TH (Xu et al. 2007; Zoeller et al. 2005). Given these data, we hypothesized that potential relationships between BPA and TH may be stronger when the time between the two measurements was shorter. We tested this hypothesis by conducting additional analyses using the BPA measurement that was closest, and the one farthest, in time to the TH measurement. For the neonatal TSH analysis, we also conducted stratified analysis by examining exposure in each trimester of pregnancy. Finally, we considered interaction by iodine intake dichotomized using the U.S. Institute of Medicine recommended daily allowance of 220 µg/day during pregnancy as the cutoff (Otten et al. 2006) based on the hypothesis that women with low iodine intake may be more susceptible to TH disruption. For neonates, we also examined effect modification by sex based on results from experimental studies (Xu et al. 2007) and by age (dichotomized at 24 hr). We set statistical significance at p < 0.05 for main effects and p < 0.10 for effect modification.

For concentrations below the LOD, we used values generated by the instrument when available. Otherwise (e.g., when no signal was detected), we imputed values at random based on a log-normal probability distribution.

### Table 1. Geometric mean (GSD) of average urinary BPA concentrations (µg/g creatinine) during pregnancy by demographic characteristics in CHAMACOS study participants.

| Characteristic | Maternal TH analyses (n = 335) | Neonatal TSH analyses (n = 364) |
|---------------|-------------------------------|---------------------------------|
|               | GM (GSD)                      | GM (GSD)                        |
| Maternal characteristics | n (%)                       | n (%)                          |
| Age (years) | 18–24  | 122 (36) | 2.0 | 3.0 | 1.9 | 1.9 |
|               | 25–29  | 132 (39) | 2.1 | 3.1 | 2.1 | 2.1 |
|               | 30–34  | 143 (42) | 2.2 | 3.2 | 2.2 | 2.2 |
|               | 35–45  | 24 (7)   | 1.5 | 2.3 | 1.7 | 2.0 |
| Race/ethnicity | Caucasian | 7 (2) | 1.6 | 2.0 | 1.7 | 2.0 |
|               | Latino   | 321 (96) | 1.3 | 2.0 | 1.9 | 2.0 |
|               | Other    | 7 (2)    | 1.6 | 1.7 | 1.5 | 1.7 |
| Education    | ≤ 6th grade | 141 (42) | 1.2 | 2.0 | 1.5 | 2.0 |
|               | 7–12th grade | 117 (35) | 1.3 | 2.1 | 2.0 | 2.0 |
|               | ≥ High school | 77 (23) | 1.3 | 1.9 | 1.5 | 1.8 |
| Family income | Poverty line | 205 (61) | 1.2 | 2.0 | 1.6 | 2.1 |
|               | Poverty line to 200% | 117 (35) | 1.3 | 2.0 | 1.6 | 2.0 |
|               | > Poverty line | 13 (4) | 1.7 | 1.6 | 1.5 | 1.6 |
| Country of birth | United States | 41 (12) | 1.4 | 2.0 | 2.0 | 2.0 |
|               | Mexico   | 287 (86) | 1.2 | 2.0 | 1.9 | 2.0 |
|               | Other    | 7 (2)    | 1.6 | 1.8 | 2.0 | 2.0 |
| Time in United States (years) | ≤ 1 | 84 (25) | 1.2 | 1.9 | 1.4 | 1.9 |
|               | 2–5     | 99 (30) | 1.2 | 2.1 | 1.8 | 1.8 |
|               | 6–10    | 77 (23) | 1.3 | 1.9 | 1.6 | 1.8 |
|               | ≥ 11    | 41 (12) | 1.4 | 2.1 | 1.6 | 2.2 |
|               | Entire life | 34 (10) | 1.5 | 2.1 | 1.6 | 2.1 |
| Parity        | 0       | 111 (33) | 1.2 | 1.9 | 1.6 | 2.0 |
|               | ≥ 1     | 224 (67) | 1.3 | 2.0 | 1.6 | 2.1 |
| Smoking during pregnancy | Yes | 23 (7) | 1.5 | 2.0 | 1.5 | 2.1 |
|               | No      | 312 (93) | 1.2 | 2.0 | 1.6 | 2.0 |
| Alcohol during pregnancy | Yes | 4 (1) | 1.1 | 1.5 | 1.6 | 1.5 |
|               | No      | 331 (99) | 1.3 | 2.0 | 1.6 | 2.0 |
| Illegal drug use during pregnancy | Yes | 5 (2) | 1.3 | 1.3 | 1.7 | 1.7 |
|               | No      | 330 (98) | 1.3 | 2.0 | 1.7 | 2.0 |
| Prepregnancy body mass index (kg/m²) | ≤ 25 | 128 (40) | 1.3 | 2.0 | 1.7 | 2.0 |
|               | 25–30   | 128 (40) | 1.2 | 1.9 | 1.7 | 1.9 |
|               | > 30    | 66 (21) | 1.3 | 2.3 | 1.7 | 2.3 |
| Infant characteristics | Sex | Male | 188 (52) | 1.3 | 1.9 | 1.7 | 1.7 |
|               | Female  | 176 (48) | 1.3 | 2.1 | 1.7 | 2.1 |
| Birth weight (g) | ≤ 2,500 | 13 (4) | 1.6 | 1.6 | 1.6 | 1.6 |
|               | 2,500–3,500 | 189 (52) | 1.2 | 1.9 | 1.6 | 1.9 |
|               | > 3,500 | 162 (45) | 1.3 | 2.2 | 1.8 | 2.2 |
| Gestational age at birth (weeks) | < 37 | 21 (6) | 1.6 | 1.9 | 1.7 | 1.9 |
|               | 37–42   | 342 (94) | 1.3 | 2.0 | 1.7 | 2.0 |
|               | > 42    | 0 (0)   | —    | —   | —    | —   |

*p < 0.05 based on analysis of variance (ANOVA).
whose parameters were determined by maximum likelihood estimation (Lubin et al. 2004). This method has been shown to yield reasonable estimates when detection frequencies are > 70%. We had complete data on most covariates. We imputed values of missing covariates (≤ 5%) at random based on observed probability distributions.

We conducted sensitivity analysis to evaluate the robustness of our results. We first re-run models after excluding outliers identified using the generalized extreme studentized deviate many-outlier procedure (Rosner 1983). We also determined whether the method that we chose to adjust for urine dilution (i.e., by dividing urinary BPA by creatinine concentration) affected our results by running models with unadjusted (with and without controlling for creatinine in models) and specific gravity–adjusted BPA concentrations. Finally, we applied inverse probability weights in an attempt to account for selection bias due to exclusion from final models (Hernan et al. 2004). We determined weights by multiple logistic regression whose independent variables were selected based on a deletion–substitution–addition algorithm (Sinisi and van der Laan 2004). Estimates from all of the above models were similar (data not shown). We present results using creatinine-adjusted BPA and including outliers in unweighted regression models.

Results

Participant characteristics. Study participants were primarily young (80% < 30 years of age) Latinas (96%) born in Mexico (84–86%) who had immigrated to the United States within 10 years of enrollment (74–78%) (Table 1). Most women had low income (60–61% below the federal poverty line) and educational attainment (77–78% did not complete high school), and many were multiracial (63–67%). Few women smoked (5–7%), consumed ≥ 1 alcoholic drink per week (1%), or used illegal drugs (1–2%) during pregnancy. There were slightly more males (52%) than female infants. Only 4% of infants were born at low birth weight (< 2,500 g), and 6% were preterm (< 37 weeks gestation). As many as 8.2% of women had iodine intakes below the recommended daily allowance.

Maternal BPA urine concentrations. BPA was detected in 82% of samples. Median BPA concentrations were similar in the first and second half of pregnancy (Table 2) as well as for BPA measurements closest and farthest in time to thyroid hormone measurements (medians = 1.1–1.2 µg/g creatinine) [see Supplemental Material, Table S1 (http://dx.doi.org/10.1289/ehp.1205092)]. BPA concentrations measured in the first and second half of pregnancy were weakly but significantly correlated with Pearson’s and intraclass correlation coefficients of 0.16 each (p < 0.01). Women who had lived their entire life in the United States (GM = 1.5 µg/g creatinine) had higher BPA urine concentrations than those who lived in the United States ≤ 1 year (GM = 1.1–1.2 µg/g creatinine) (Table 1). Median BPA concentrations in our study population were lower than in women of reproductive age (18–44 years) in the 2007–2008 NHANES sample (median = 1.9 µg/g creatinine; IQR = 1.2–3.4) (CDC 2011).

Maternal and neonatal TH levels. Mean ± SD serum concentrations of free and total T4 were 0.8 ± 0.2 ng/dL and 10.6 ± 1.6 µg/dL in maternal samples, respectively. The GM was 1.2 mIU/L (GSD = 1.7) for maternal TSH and 5.6 mIU/L (GSD = 1.8) for neonatal TSH. Based on trimester-specific laboratory reference ranges, four women had elevated free T4 (> 1.6 µg/dL) and 39 had low TSH (< 0.5 and < 0.8 mIU/L in second and third trimesters, respectively); none had high total T4 (> 1.78 and > 20.1 µg/dL in the second and third trimester, respectively). Seven women had low free T4 (< 0.5 ng/dL), 13 had low total T4 (< 8.0 µg/dL) and two had elevated TSH (> 4.6 and > 5.2 mIU/L in the second and third trimester, respectively). Seventeen had high TSH based on National Academy of Clinical Biochemistry guidelines (> 2.5 mIU/L) (Mandel et al. 2005). TSH was elevated (> 25 mIU/L) in one of the neonates.

Association between maternal BPA urine concentrations and maternal and neonatal TH levels. Maternal urinary BPA concentrations were not associated with maternal free T4 or TSH, but were negatively associated with maternal total T4 (Table 3). This association was not statistically significant using the average of the two BPA measurements (β = −0.13; 95% CI: −0.29, 0.02) or using the BPA measurements taken farthest in time (median = 95 days; IQR = 63–116) to total T4 serum measurements (β = −0.06; 95% CI: −0.20, 0.08), but was significant when the BPA measurements taken closest in time to total T4 measurements (median = 6 days; IQR = 0–15) were examined (β = −0.13; 95% CI: −0.25, 0.00). Average maternal urinary BPA was also related to increased odds of low total T4 [odds ratio (OR) = 1.6; 95% CI: 1.1, 2.3] and low TSH (OR = 1.5; 95% CI: 1.1, 2.0). Similar to results from linear regressions, associations were stronger when the

| Table 2. Urinary BPA concentrations (µg/g creatinine) during pregnancy in CHAMACOS study participants in samples included in analyses of maternal (n = 335) and neonatal (n = 364) TH hormone levels.
| Timing of measurement | Percentile | Detection frequency (%) |
|-----------------------|------------|------------------------|
| Maternal TH analyses  |            |                        |
| First half of pregnancy | 290        | 0.6                    |
| Second half of pregnancy | 317        | 0.7                    |
| Pregnancy average     | 335        | 0.8                    |
| Neonatal TSH analyses |            |                        |
| First half of pregnancy | 309        | 0.7                    |
| Second half of pregnancy | 344        | 0.7                    |
| Pregnancy average     | 364        | 0.8                    |

| Table 3. Associations between maternal BPA urinary concentrations during pregnancy and TH levels in women and their neonates participating in the CHAMACOS study.
| Outcome | Exposure time | Unadjusted models | Adjusted models |
|---------|---------------|-------------------|----------------|
| Maternal TH analyses | | | |
| Free T4 | Closest measurement | 332 | 0.00 (-0.02, 0.02) | 0.00 (-0.02, 0.02) |
| Free T4 | Farthest measurement | 332 | 0.01 (-0.01, 0.03) | 0.01 (-0.01, 0.03) |
| Free T4 | Pregnancy average | 332 | 0.00 (-0.02, 0.03) | 0.00 (-0.02, 0.03) |
| Total T4 | Closest measurement | 335 | -0.11 (-0.25, 0.02) | -0.13 (-0.25, 0.00)* |
| Total T4 | Farthest measurement | 335 | -0.05 (-0.18, 0.09) | -0.06 (-0.20, 0.08) |
| Total T4 | Pregnancy average | 335 | -0.12 (-0.28, 0.04) | -0.13 (-0.28, 0.02) |
| TSH* | Closest measurement | 335 | -2.9 (-7.5, 2.0) | -3.5 (-8.2, 1.5) |
| TSH* | Farthest measurement | 335 | 0.3 (-4.7, 5.4) | 0.1 (-4.9, 5.5) |
| TSH* | Pregnancy average | 335 | -2.5 (-8.2, 3.6) | -3.3 (-8.2, 2.3) |
| Neonatal TSH analyses | | | |
| TSH* | Closest measurement | 364 | -2.4 (-7.3, 2.7) | -2.0 (-6.1, 2.2) |
| TSH* | Farthest measurement | 364 | -1.2 (-4.7, 5.4) | 0.5 (-5.2, 3.3) |
| TSH* | Pregnancy average | 364 | -1.7 (-8.1, 5.2) | -1.8 (-4.2, 0.8) |

Coefficients represent the mean change (free and total T4) or percent change (TSH) in thyroid hormone levels for each doubling in maternal BPA urinary concentrations. Maternal TH models were adjusted for mother’s age, education level, country of birth, poverty level, alcohol and drug use during pregnancy, iodine intake, and hexachlorobenzene and polychlorinated biphenyl serum concentrations. Neonatal TSH models were adjusted for mother’s country of birth and child’s age at TSH measurement.  

*Percent change in TSH serum concentration calculated using the following formula: (10^p – 1) * 100. *p < 0.05.

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BPA and TH measurements were taken closer together relative to when measurements were taken farther apart (data not shown).

Although maternal BPA urinary concentrations were not associated with neonatal TSH when both sexes were combined (Table 3), effect modification by sex was statistically significant ($p = 0.01$). Analyses stratified by sex revealed an inverse relationship among boys ($-9.9\%$ for every doubling in average BPA; $95\%$ CI: $-15.9\%$, $-3.5\%$) but not girls ($4.4\%$ for every doubling in average BPA; $95\%$ CI: $-2.4\%$, $11.7\%$) (Figure 1). Among boys, associations with neonatal TSH were stronger when maternal BPA urinary concentrations were measured in the third trimester of gestation ($-9.3\%$ for every doubling in BPA; $95\%$ CI: $-16.3\%$, $-1.7\%$) than when BPA was measured in the first ($-3.2\%$; $95\%$ CI: $-11.5\%$, $6.0\%$) or second ($-5.1\%$; $95\%$ CI: $-11.3\%$, $1.6\%$) trimesters. Associations were weaker as the time between the BPA and TSH measurements increased. Results were similar to those obtained using third trimester BPA when analyses were restricted to the BPA samples collected closest in time to the TSH measurements (median = 92 days; IQR = 81–104) (data not shown). None of the above associations were significantly modified by iodine intake or age at the time of heel stick (data not shown).

There were no statistically significant differences in demographic or lifestyle variables between participants included in maternal TH analyses relative to those excluded. Participants who were included in neonatal TSH analyses were more educated (22% ≥ high school education vs. 16%), were less likely to have used illegal drugs during pregnancy (2% vs. 5%), and delivered children with a higher birth weight (mean = 3,438 g vs. 3,314 g) ($p < 0.05$).

**Discussion**

We report significant inverse associations between maternal BPA urine concentrations during pregnancy and TSH in male, but not female, neonates after adjustment for covariates. The relationship among males was stronger when BPA was measured in the third trimester of gestation. We also found an inverse association between maternal urinary BPA and serum total $T_4$ when analyses were restricted to the BPA measurement taken closest in time to the TH measurement. However, contrary to a previous study investigating relations between urinary perchlorate and serum TSH and $T_4$ using NHANES data (Blount et al. 2006), we found no evidence of a stronger relation among women with low iodine intake.

This is the first study to evaluate associations between maternal BPA urine concentrations during pregnancy and maternal or neonatal TH in humans. Meeker and Ferguson (2011) reported a borderline significant ($p = 0.08$) inverse association between total $T_4$ in serum and BPA concentrations in the urine of 1,367 adult NHANES participants, consistent with our findings. However, no association was found with free $T_4$ or total $T_3$ (total $T_4$ was not measured) in a smaller study of 167 men from an infertility clinic, but an inverse relation with TSH was reported (Meeker et al. 2010). On the other hand, BPA urinary concentrations were not found to be related to hypothyroidism in 45 Japanese women with a history of miscarriage (Sugiuara-Ogasawara et al. 2005). In addition to sample size, inconsistent results may be attributable to differences in study populations. Whereas one study was conducted in a nationally representative sample (Meeker and Ferguson 2011), other investigations examined the question in individuals who may have suffered from medical conditions (Meeker et al. 2010; Sugiuara-Ogasawara et al. 2005). In addition, prior studies examined relations between BPA and TH among nonpregnant adults only. Fetuses and pregnant women may however be particularly susceptible to BPA exposure. Livers of pregnant rats have been shown to have a lower capacity to glucuronidate BPA relative to nonpregnant rats (Inoue et al. 2005), and the human fetus has limited glucuronidation capacity (Ring et al. 1999).

Of interest is the apparent sexually dimorphic relationship of BPA and neonatal TSH, suggesting an association in males but not females. Although data from NHANES suggested that inverse relations between BPA and total $T_4$ were similar for both sexes in adults (Meeker and Ferguson 2011), one experimental rat study found that exposure to BPA during pregnancy and postpartum was related to alterations in free $T_4$ (total $T_4$ was not measured) among male pups only (Xu et al. 2007). Furthermore, results from some animal studies suggest that males may not metabolize BPA as efficiently as females. For instance, mRNA expression of the uridinediphosphate glucuronol transferase (UDP-GT) isomor UGT 2B1 [which glucuronidates BPA in rats (Yokota et al. 1999)] has been reported to be lower in male rats than in females (Takeuchi et al. 2004). In addition, exposure to BPA was found to lower the expression of UGT 2B1 in male rats but not in females (Shibata et al. 2002).

BPA is eliminated quickly in humans (half-life < 6 hr) (Volkel et al. 2002), and data from our study as well as others’ (Braun et al. 2011; Ye et al. 2011), which show that the between-person variance of creatinine-adjusted BPA measurements accounted for only 9–16% of the total variance, suggest that a large number of measurements may be necessary to assess exposure over the long term. Urinary BPA measurements may, however, be adequate to evaluate transitory effects. Some experimental studies suggest that exposure to BPA may have such effects on TH (Xu et al. 2007; Zoeller et al. 2005). Given these results, we hypothesized that relations between BPA and TH may be stronger when the time elapsed between the two measurements was shorter. Our finding that the association between urinary BPA and maternal serum total $T_4$ was significant when BPA was measured closer in time, but not when measurements were farther apart, and that the relation with neonatal TSH was stronger when BPA was measured in the third trimester of pregnancy appears to support this hypothesis or may be attributable to a specific developmental window of susceptibility to BPA. It is noteworthy that effect estimates for the average BPA and the closer measurements were very similar albeit of slightly different precision. Average urinary BPA may therefore remain a useful measure in studies of thyroid hormone disruption.

BPA may affect thyroid function through a number of pathways. In addition to binding and activating the TR (Moriyama et al. 2002), BPA has been shown to induce UDP-GT activity in European polecats (Nieminen et al. 2002). Since UDP-GT regulates the rate-limiting step in $T_4$ metabolism in rats and presumably in humans, induction of this enzyme could underlie the reduction in $T_4$ observed in this study. However, contrary to some hydroxylated metabolites of other commonly measured environmental chemicals, such as PCBs and PBDEs, BPA binds only weakly to the transport proteins transthyretin and thyroxine-binding globulin (Marchesini et al. 2008). The finding of a reduction in total $T_4$ but not free $T_4$ was unexpected. This may be attributable to a BPA-induced reduction in the serum concentration of transport proteins, an increased elimination of free $T_4$ compensated by a release of $T_4$ from transport proteins, or a hepatic sequestration of $T_4$ as observed following exposure of mice to PCB-153 (Kato et al. 2011).
Strengths of this study include the availability of data on a large number of potential confounders including iodine intake during pregnancy. Iodine is an essential component of TH and both iodine deficiency and excess can adversely affect thyroid function (Delange et al. 2002). Data from NHANES suggest that, although the United States is generally considered to have an iodine sufficient area, a significant proportion of pregnant U.S. women are iodine deficient (Caldwell et al. 2008). We also conducted sensitivity analysis and our results were robust to methods used to adjust for urine dilution (creatinine or specific-gravity adjustment) and to adjustment for selection bias due to exclusion from final models.

This study also has some limitations. Associations that we report were identified in an immigrant Mexican-American population with low socioeconomic status. Our results thus need to be confirmed in other populations to evaluate their generalizability. Although we considered many variables known to affect TH levels, unmeasured confounders may have affected our results. In addition, the health implications of a possible decrease in maternal T₃ with no reduction in free T₃ is unclear because bound T₃ is not biologically active (Mendel 1989). Similarly, we are aware of no studies that investigated the developmental effect of lower but normal neonatal T₃.

In summary, we report an inverse relation between BPA concentration in maternal urine and maternal serum total T₄ during pregnancy. Although we cannot rule out that average BPA concentrations during pregnancy may be relevant, the association of maternal BPA and total T₄ was stronger when they were measured closer together relative to further apart in time, suggesting a transient effect of BPA. Similarly, the relationship of maternal BPA and neonatal T₃ was strongest when maternal BPA was measured in the third trimester compared with earlier in pregnancy. This may also suggest a transient effect of BPA on maternal thyroid hormone production. In conclusion, the health implications of a possible decrease in maternal T₃ with no reduction in free T₃ is unclear because bound T₃ is not biologically active (Mendel 1989).

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