RESEARCH ARTICLE

Population Structure of *Montastraea cavernosa* on Shallow versus Mesophotic Reefs in Bermuda

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Abstract

Mesophotic coral reef ecosystems remain largely unexplored with only limited information available on taxonomic composition, abundance and distribution. Yet, mesophotic reefs may serve as potential refugia for shallow-water species and thus understanding biodiversity, ecology and connectivity of deep reef communities is integral for resource management and conservation. The Caribbean coral, *Montastraea cavernosa*, is considered a depth generalist and is commonly found at mesophotic depths. We surveyed abundance and size-frequency of *M. cavernosa* populations at six shallow (10m) and six upper mesophotic (45m) sites in Bermuda and found population structure was depth dependent. The mean surface area of colonies at mesophotic sites was significantly smaller than at shallow sites, suggesting that growth rates and maximum colony surface area are limited on mesophotic reefs. Colony density was significantly higher at mesophotic sites, however, resulting in equal contributions to overall percent cover. Size-frequency distributions between shallow and mesophotic sites were also significantly different with populations at mesophotic reefs skewed towards smaller individuals. Overall, the results of this study provide valuable baseline data on population structure, which indicate that the mesophotic reefs of Bermuda support an established population of *M. cavernosa*.

Introduction

In recent years, coral reefs have undergone drastic decline due to numerous anthropogenic impacts to environmental conditions including eutrophication, disease, the loss of herbivory, and bleaching associated with ocean warming [1–4]. Currently, nearly 30% of the world’s coral reefs are considered severely damaged, and close to 60% are in danger of being lost by 2030 [5]. These losses are particularly pronounced on shallow water reefs of the Caribbean, where the comprehensive study by Jackson et al. [1] reports an overall decline in coral cover of 59%, from an average of 33% before 1984 to 14.3% since 2005. Deep reef systems in the mesophotic zone (>30m), however, have not experienced the same trend, displaying relatively stable coral
populations over time [6]. Yet, in comparison to shallow-water coral reefs, mesophotic reefs have received little attention [7].

Mesophotic coral ecosystems (MCE’s) are comprised of a variety of taxa, including sponges, macroalgae, and azooxanthellate corals, as well as light-dependent zooxanthellate corals that exist in zones between approximately 30m and 150m in tropical and subtropical zones [8–10]. These regions tend to exist in low energy deep fore-reef zones that are characterized by steep gradients in light and temperature [8]. Typically the depth at which light is reduced to 1% of the available surface light defines the lower limits of the mesophotic zone [11]. Previous technological limitations have presented major challenges to conducting research on MCE’s, resulting in limited understanding of the bathymetric and geographic extent of MCE’s and the biodiversity and community structure they support across regions. Even basic taxonomic and systematic characterization of these communities is unknown, underscoring the importance of establishing baseline information on species assemblages and the roles they play in ecosystem function [10, 12].

Analyses of size-frequency distributions can reveal characteristics of species populations as they represent stages of population growth and decline [13]. Population size structure results from variations in rates of colony growth, recruitment and mortality, and may indicate individual sensitivities to life-history processes and environmental variation. The life cycle of modular organisms such as scleractinian corals, however, is complicated by processes such as fragmentation, fission, fusion, and partial mortality, making the relationship between surface area and age difficult to interpret [14]. Yet coral colony surface area can be correlated to age if partial mortality is low, and thus characterizations of population size-frequency distributions may provide critical demographic information, particularly for massive, non-branching colonies [13, 15–19]. Describing coral populations in terms of population size-frequency, therefore, can provide a snap-shot of current reef condition and if monitored over time may serve as an indicator for stability or decline [17, 20].

The aim of this study is to provide baseline characterization of population structure for the dominant zooxanthellate coral at adjacent shallow and mesophotic reefs in Bermuda. As such, this study provides an initial assessment of mesophotic reef condition in relation to environmental conditions that vary with depth, such as temperature and nutrient levels. Fricke and Meischner [21] conducted the only comprehensive study of mesophotic reef composition in Bermuda using submersible video transect surveys. Their study found species diversity decreased drastically below 40m. Among the species found in these mesophotic zones include Agaricia fragilis, Stephanocoenia michelini, Madracis decactis, Scolymia cubensis, Montastraea cavernosa and Orbicella franksii, with M. cavernosa and O. franksii being the dominant representatives below 30m. Using in situ diver-led surveys we examine variations in colony density, surface area, percent cover and size-frequency distributions and provide baseline data on the population structure of M. cavernosa on mesophotic reefs in Bermuda.

**Materials and Methods**

**Ethics Statement**

Surveys for this study were conducted in public areas outside of any marine reserves and did not require approval or permitting. No specimens were manipulated or collected from reef sites in completing this study and care was taken to avoid contact with benthic substrata.

**Site and Species Selection**

Located 32°N, 64°W, Bermuda’s sub-tropical coral reefs represent the northernmost reef system in the Atlantic. The shallow rim reefs of this pseudo-atoll encircle the platform, dropping
quickly to deep mesophotic reefs. Thus, deep reefs are easily accessible in Bermuda and corals surviving in these zones are both at their latitudinal and bathymetric limits. Furthermore, shallow water coral cover in Bermuda ranks among the highest in the Caribbean with an estimated cover of 38.6% [1]. Bermuda, therefore, is an ideal and important location in which to study coral community composition and connectivity across a depth gradient.

*M. cavernosa* (Linnaeus, 1767) is a common reef building coral on fore reef slopes throughout the Caribbean and western Atlantic, extending from Bermuda to Brazil and the West African coast [22, 23]. *M. cavernosa* is considered an ‘extreme’ depth-generalist [9], as it inhabits depths from 3–100m across its geographical range [21, 24, 25]. Along its bathymetric distribution, *M. cavernosa* exhibits significant phenotypic plasticity in morphology, rates of respiration, and primary productivity [25–27], and is the only hermatypic species documented to survive below 70m in Bermuda [21].

**Surveys**

Twelve coral surveys were performed between August 17th and December 28th 2014 to estimate abundance and surface area of *M. cavernosa* colonies between shallow and mesophotic reef sites in Bermuda (Fig 1). Six surveys were conducted at shallow sites (10m depth), and six were conducted at nearby mesophotic sites (45m depth). Site names, map labels, GPS coordinates, and survey dates are included in Table 1. Site locations were selected based on accessibility and

![Survey Map](https://example.com/figure1.png)
visual identification of reef structure at mesophotic sites. Paired shallow sites were selected as the nearest site encountered at 10m depth traveling up the reef slope perpendicular to the shoreline. At each site, all M. cavernosa colonies with greater than 50% of the colony located within 1m of either side of a 30m transect tape (60m² total area per survey) were counted and largest surface diameter measured (to nearest cm). Diameter was chosen as a metric for ease of completing surveys at depth with minimal bottom time (maximum bottom time of 25 min.). Transects at each site were laid along the reef slope to ensure a constant depth, beginning at the closest non-living reef structure encountered upon reaching the benthos to which the tape could be secured. A colony was defined as any autonomous coral skeleton with living tissue as described by Meesters et al. [19].

### Population Structure Analyses

Density of M. cavernosa colonies (# of colonies 60m⁻²) met the assumptions of normality and equal variance and was analyzed by depth using a Student’s t-test. Mean colony diameter was used to calculate surface area of each colony using the following equation: surface area = 2π (diameter/2)². Colony surface area was logarithmically transformed to reduce non-normality and heteroskedasticity and for each site, geometric mean (μ), standard deviation (SD), skewness (g1), and kurtosis (g2) were calculated. Mean colony surface area, standard deviation and skewness were compared by depth using the Student’s t-test (n = 6), and kurtosis was compared by depth using a Mann-Whitney U-test [28–30]. These statistics describe the shape of a distribution and allow comparisons between populations at different depths independent of colony surface area [13, 19]. The total surface area per 60m² transect was also used to calculate percent cover of M. cavernosa (% cover 60m⁻²). Data was transformed to arcsine values and compared by depth using a Student’s t-test.

Mean size-frequency distributions were generated for each depth zone (shallow and mesophotic) and compared with each other by a Kolmogorov-Smirnov test and to a normal distribution using a Shapiro-Wilk W test [31–33]. Additionally, size-frequency distributions within each shallow and mesophotic site were compared using a Kolmogorov-Smirnov test. Similarity of size-frequency distributions between shallow and mesophotic sites was calculated with the Spearman rank-correlation coefficient by dividing colony numbers into 10 surface area size classes based on a logarithmic scale (class borders were < 0.5, 1.0, 1.5, 2.0, 2.5, 3.0, 3.5, 4.0, 4.5,

| Map Label | Site Name | Depth (m) | Date       | Latitude   | Longitude |
|-----------|-----------|-----------|------------|------------|-----------|
| S1        | Rita      | 10        | 17-Aug-14  | N32° 21' 29.3" | W64° 38' 29.3" |
| S2        | Coopers   | 10        | 17-Aug-14  | N32° 20' 28.4" | W64° 39' 28.1" |
| S3        | Tuckers   | 10        | 5-Sep-14   | N32° 19' 57.7" | W64° 40' 16.5" |
| S4        | Spittal   | 10        | 5-Sep-14   | N32° 18' 42.3" | W64° 42' 53.4" |
| S5        | Devonshire| 10        | 28-Dec-14  | N32° 18' 0.7"  | W64° 44' 16.1" |
| S6        | Hungry Bay| 10        | 28-Dec-14  | N32° 17' 8.5"  | W64° 45' 26.1" |
| D1        | XL        | 45        | 17-Aug-14  | N32° 21' 58.0" | W64° 36' 5.3"  |
| D2        | Coopers   | 45        | 17-Aug-14  | N32° 20' 29.6" | W64° 37' 47.2" |
| D3        | Tuckers   | 45        | 5-Sep-14   | N32° 19' 8.8"  | W64° 39' 41.0" |
| D4        | Spittal   | 45        | 5-Sep-14   | N32° 18' 3.7"  | W64° 42' 34.7" |
| D5        | Devonshire| 45        | 21-Dec-14  | N32° 17' 36.5" | W64° 43' 48.9" |
| D6        | Hungry Bay| 45        | 28-Dec-14  | N32° 16' 37.5" | W64° 44' 39.4" |

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4.5 cm²). Correlation coefficients were not normally distributed, and group means were tested with the Mann-Whitney U-test. All analyses were computed using PASW Statistics 17.0.

Size-frequency distributions within sites were also examined with a principal coordinate ordination (PCO) analysis based on Euclidean similarity, which generates a two-dimensional plot. PCO analysis is an equivalent to principal component analysis (PCA), but with more flexibility of resemblance measures [34] and allows spatial visualization of dissimilarities among sites and between depths. This analysis was performed using PRIMER version 6.

**Nutrient and Temperature Analyses**

Seawater samples were collected at two shallow sites and two mesophotic sites during survey dives (Tuckers and Spittal). Four replicate samples were collected at each site. Analysis of nitrate (NO₃), nitrite (NO₂), and silicate (SiO₄²⁻) were conducted at BIOS with a Seal Analytical AA3 continuous flow analyzer. Concentrations of nitrogen (NO₃ + NO₂) and silicate at each site met the assumptions of normality and equal variance and were analyzed by depth using Student’s t-tests (n = 4 per site). Seawater temperature readings were recorded at each of the surveyed shallow and mesophotic sites between July 2014 and January 2015 using a Shearwater Petrel dive computer. Each site was visited twice during this time period for a total of 12 paired temperature readings. Mean temperatures were compared by depth using a Student’s t-test (n = 6).

**Results**

**Distribution Parameters**

Table 2 gives the geometric mean surface area, skewness, kurtosis, maximum colony surface area, standard deviation, the probability that the sample is from a normal distribution, and the sample size at each site. Each parameter is also given for all shallow sites and all mesophotic sites combined.

| Site    | Depth | μ     | g₁    | g₂    | SD   | 95%   | P_norm | n   |
|---------|-------|-------|-------|-------|------|-------|--------|-----|
| Rita/XL | 10    | 2508  | -0.638| 0.497 | 0.834| 16343 | 0.042  | 26  |
| Coopers | 10    | 1731  | 0.389 | -0.369| 0.518| 8836  | 0.012  | 12  |
| Spittal | 10    | 1503  | -0.535| -0.281| 0.457| 5655  | 0.006  | 21  |
| Tuckers | 10    | 1991  | -1.904| 4.283 | 0.737| 5284  | 0.000  | 6   |
| Devonshire | 10   | 1808  | -0.154| -1.161| 0.555| 7697  | 0.005  | 29  |
| Hungry Bay | 10 | 2145  | 0.046 | -1.052| 0.617| 13586 | 0.010  | 31  |
| Rita/XL | 45    | 522   | -0.755| 0.992 | 0.687| 7697  | 0.009  | 62  |
| Coopers | 45    | 330   | -0.763| 0.630 | 0.555| 2389  | 0.010  | 66  |
| Spittal | 45    | 639   | -0.304| -0.178| 0.709| 3927  | 0.043  | 36  |
| Tuckers | 45    | 349   | -1.088| 0.945 | 0.767| 3041  | 0.031  | 58  |
| Devonshire | 45 | 322   | -0.260| 0.004 | 0.749| 2513  | 0.152  | 96  |
| Hungry Bay | 45 | 441   | -0.614| 0.273 | 0.719| 2513  | 0.078  | 108 |
| Shallow | 10    | 1933  | -0.499| 0.452 | 0.620| 16343 | 0.000  | 125 |
| Deep   | 45    | 434   | -0.627| 0.369 | 0.713| 7697  | 0.000  | 426 |

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The mean density of colonies varied significantly by depth (p = 0.002, Students t-test, F = 0.106, n = 6), with higher colony density at mesophotic sites compared with shallow sites.
Colony surface area also varied significantly between depths ($p < 0.0001$, Students’ $t$-test, $F = 0.082$, $n = 6$), where mean colony surface area was smaller at mesophotic sites compared with shallow sites (Fig 2B; Table 2). Mean colony surface area at shallow sites was typically 4.5 times greater than at deeper sites, and maximum surface area was 2.1 times greater at shallow sites (16343 cm$^2$) compared with mesophotic sites (7697 cm$^2$). This large discrepancy in individual colony surface area resulted in relatively equal contributions to mean percent cover at each depth ($p = 0.322$, Students’ $t$-test, $F = 0.091$, $n = 6$), despite the higher density of colonies at mesophotic sites (Fig 2C).

### Standard deviation, skewness and kurtosis

Standard deviations of colony surface area data did not differ significantly between shallow and mesophotic sites ($p = 0.262$, Student’s $t$-test). This suggests that variation in colony surface area is similar at shallow and mesophotic sites.

The asymmetry around the mean of a size-frequency distribution is described as the skewness ($g_1$; Table 2); where a negative $g_1$ describes a distribution skewed to the left and a positive $g_1$ distribution is skewed to the right. In a perfectly symmetrical distribution, $g_1$ is zero [19]. Skewness did not vary significantly by depth ($p = 0.649$, Student’s $t$-test). Distributions at mesophotic and shallow sites were negatively skewed, indicating a lower frequency of colonies in the smaller size classes.

The degree of peakedness of a distribution around its central mean is described as kurtosis ($g_2$), where a population can be either over centralized (leptokurtic, $g_2 > 0$) or flatter than normal (platykurtic, $g_2 < 0$). Kurtosis did not vary significantly by depth ($p = 0.150$, Mann-Whitney $U$-test), where the average kurtosis was 0.45 and 0.37 for shallow and mesophotic sites, respectively.

### Size-Frequency Distributions

Mean size-frequency distributions for shallow versus mesophotic sites are given in Fig 4. Logarithmically transforming colony surface area data greatly improved normality. Mean distribution patterns from shallow and mesophotic sites were bell-shaped, yet differed significantly from a normal distribution ($p < 0.05$, Shapiro-Wilk W test). Furthermore, mean distribution differed significantly between shallow versus mesophotic sites, being skewed towards larger colonies at shallow sites compared with mesophotic sites ($p < 0.001$, Kolmogorov-Smirnov test).

Distributions within each of the mesophotic sites were bell-shaped, and 2 out of six sites did not differ from normal distribution ($p > 0.05$). Distributions within the shallow sites were more variable due to the lower density of individuals, with distributions at all sites differing from a normal distribution ($p < 0.05$). Similarity of size-frequency distributions from each site were compared using the Spearman rank-correlation coefficient. These comparisons showed that distributions from the same depths (from distant sites) were more similar than those from adjacent sites at different depths (Table 3). The mean correlation coefficient of distributions from sites at the same depths was 0.29 (SD = 0.08, $n = 30$), while the mean correlation coefficient of comparisons from adjacent sites at different depths was 0.24 (SD = 0.28, $n = 6$). These means are significantly different ($p = 0.006$, Mann-Whitney $U$-test). The high degree of similarity between distributions from the same depth suggests that the population structure of *M. cavernosa* has depth specific characteristics. The PCO results are provided in Fig 6, confirming a clear separation of the size-frequency distributions between depths and more similarity among sites of the same depth than between paired sites at different depths.
Fig 2. Colony Abundance, Surface Area, and Percent Cover by Depth. (a) mean number of *M. cavernosa* colonies per 60m² ± SE at shallow (10m; gray bars) versus mesophotic (45m; black bars) sites (Rita/XL, Coopers, Tuckers, Spittal, Devonshire, Hungry Bay); (b) mean *M. cavernosa* colony surface area (cm²) ± SE at shallow (10m; gray bars) versus mesophotic (45m; black bars) sites; (c) mean percent cover ± SE of *M. cavernosa* at shallow (10m; gray bars) versus mesophotic (45m; black bars) sites (n = 6 per depth).

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Nutrient concentrations were higher on shallow sites compared with mesophotic sites (Fig 7A), with significant differences found in concentrations of nitrate and nitrite between depths \( (p < 0.0001, \text{Tuckers}; p = 0.019, \text{Spittal}; \text{Student's t}-test, n = 4) \) and silicate between depths at Tuckers \( (p = 0.001, \text{Student's t}-test, n = 4) \), but not at Spittal \( (p = 0.058, \text{Student's t}-test, n = 4) \).

Mean seawater temperatures also differed significantly by depth, being higher on shallow sites compared with mesophotic sites \( (p < 0.0001, \text{Student's t}-test, n = 6) \). Likewise, variation in temperature was more pronounced on shallow sites, ranging from 22.8 to 29.5°C, compared with mesophotic sites, ranging from 22.2 to 27.8°C (Fig 7B).

**Discussion**

This study documents the population structure of *M. cavernosa* at mesophotic versus shallow reefs in Bermuda and reveals depth specific characteristics of these populations. Our analyses show that size-frequency distributions of populations at shallow reefs vary significantly from those at mesophotic reefs (Fig 4), with colonies from neighboring reefs at the same depths being more similar to one another than to those from adjacent populations at different depths.
These results suggest that conditions that vary with depth, such as light, seawater temperature, and nutrient concentration, likely influence *M. cavernosa* population structure. Overall, this study found the distribution of populations at the mesophotic reef sites examined on Bermuda’s south shore is shifted towards smaller individuals relative to shallow reefs (Fig 4). Likewise, the average colony surface area at these mesophotic reefs was significantly smaller than at shallow reefs (Fig 2A). However, it is important to note that a large degree of variation in colony morphology was observed at the different depths, with colonies at mesophotic sites being predominately flat and disc shaped, compared with colonies at shallow sites that varied from flat, to encrusting, to massive boulders. Thus, using diameter to estimate surface area likely underestimated total surface area of shallow-water colonies, resulting in a conservative estimate of colony size differential by depth in this study. These data indicate, therefore, that growth and maximum colony surface area may be limited on mesophotic reefs and suggests that maximum *M. cavernosa* colony surface area is likely controlled by environmental conditions that may limit energetic resources, such as light and nutrient availability.
Fig 5. Size-Frequency by Site. Size-frequency distributions of *M. cavernosa* on a logarithmic scale represented as the number of individuals within each log transformed size class for colonies from each population.
The smaller surface areas of *M. cavernosa* colonies found at mesophotic reefs in this study conforms to previous studies that document a decrease in coral colony surface area with depth [35] and the predominance of small colonies at deeper depth distributions [36, 37]. Smaller colony surface area may be a result of nutrient limitation at mesophotic reefs as nutrient analyses of adjacent mesophotic and shallow reefs in this study indicate that nutrients are significantly reduced at these mesophotic sites compared with shallow sites (Fig 7A). Likewise, Lesser et al. [25] document a reduction in phytoplankton availability and marked decreases in light-dependent productivity with depth, indicating that energy required for calcification and growth may indeed be limited for mesophotic corals [38, 39]. Under low-light and nutrient-limited conditions such as those present at mesophotic reefs, corals decrease metabolic demand through reduced respiration [40], slower growth, and morphological adaptations. For example, Grigg [41] found skeletal extension rates of *Porites lobata* declined exponentially with PAR from 3 to 50m in Hawaii. Likewise, Fricke et al. [35] report skeletal extension rates in *Leptoseris fragilis* at 90 to 120m of 0.5–0.8mm year⁻¹, which is significantly lower than typical rates reported for other non-branching shallow water corals ranging from 1.0–8.5mm year⁻¹ [42]. Alternatively, Bongaerts et al. [43] report an average growth rate of 22.0mm year⁻¹ for *Agaricia grahamae* fragments transplanted to 60m in Curacao, which is similar to growth rates in the congeneric species *A. humilis* and *A. agaricites* from shallow reefs (<30m) in the same region [44, 45]. Additionally, metabolic demands may be met through increased reliance on heterotrophy in conditions where primary production is limited. In the Bahamas, Lesser et al. [25] document a transition from autotrophy to heterotrophy with depth in populations of *M. cavernosa* between 45 and 61m associated with significant declines in primary productivity. Whether the energy consumed through heterotrophy is substantial enough to compensate for the reduction in primary production and maintain metabolic rates similar to shallow corals, however, is unclear. Thus, future studies of *M. cavernosa* on mesophotic reefs should include examinations of skeletal extension rates to determine rates of growth.

Despite the smaller surface area per colony of *M. cavernosa* at mesophotic sites, the relatively high density resulted in equal contributions to percent cover at mesophotic and shallow reefs (Fig 2). The high density of *M. cavernosa* colonies at mesophotic reefs may be related to lack of competition with other coral species that are unable to adapt to conditions at this depth.

Table 3. Correlation Coefficients. Spearman rank correlation coefficient values for comparisons of size-frequency distributions of *M. cavernosa* between sites. Values above the staggered line are comparisons among shallow sites; values below the staggered line are comparisons among mesophotic sites; values between the staggered lines are between adjacent shallow and mesophotic sites. Significant correlations (statistically similar; α = 0.05) are indicated in bold.

| Spearman rank-correlation coefficient | Rita/XL | Coopers | Spittal | Tuckers | Devonshire | HungryBay |
|--------------------------------------|---------|---------|---------|---------|------------|-----------|
| Rita/XL                              | 0.641   | 0.717   | 0.279   | 0.795   | 0.653      | 0.873     |
| Coopers                              | 0.934   | 0.504   | 0.203   | 0.934   | 0.988      | 0.880     |
| Spittal                              | 0.962   | 0.881   | 0.756   | 0.418   | 0.229      | 0.306     |
| Tuckers                              | 0.856   | 0.879   | 0.715   | -0.160  | 0.903      | 0.943     |
| Devonshire                           | 0.905   | 0.931   | 0.816   | 0.892   | 0.296      | 0.847     |
| HungryBay                            | 0.917   | 0.919   | 0.829   | 0.941   | 0.939      | 0.332     |

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Diversity decreases dramatically at mesophotic reefs with only a handful of scleractinian species known to reside at these depths in Bermuda, including *A. fragilis*, *M. carambi*, *M. decactis*, *M. cavernosa*, *O. franksii*, *P. porites*, *S. michelini*, and *S. cubensis*. Among them, *M. cavernosa* is

![Fig 6. PCO of Population Structure by Site.](image)

Principal coordinates analysis (PCO) of *M. cavernosa* size-frequency distributions for each survey location (Rita/XL, Coopers, Tuckers, Spittal, Devonshire, Hungry Bay) at 10m and 45m depths. PCO1 and PCO2 axes together capture 94.7% of the total variation in size-frequency distribution.

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![Fig 7. Nutrient Concentration and Temperature by Depth.](image)

(a) mean (±SD) concentration (μM) of nitrate (NO₃) + nitrite (NO₂) and silicate (SiO₄²⁻) on shallow (10m; n = 4 per site) versus mesophotic sites (45m; n = 4 per site) from water samples collected September 5, 2015 (NO₃ + NO₂, p<0.0001, Tuckers, p = 0.019, Spittal; SiO₄²⁻, p = 0.001, Tuckers, p = 0.058, Spittal; Student's t-tests); (b) box blot of seawater temperature at shallow (10m) and mesophotic (45m) sites showing median values (solid horizontal line), 25th and 75th percentile values (box outline), and minimum and maximum values (whiskers) recorded between July 2014 and January 2015 from 6 paired shallow (10m) and mesophotic (45m) survey sites (2 dives per site); Rita/XL, Coopers, Tuckers, Spittal, Devonshire, and Hungry Bay (p<0.0001, Students t-test, n = 6).

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the most predominant and is the only coral documented to survive in Bermuda below 70m [21, 46]. On surveys conducted for this study, *M. cavernosa* was the most abundant scleractinian species at these mesophotic sites, while *A. fragilis* was the second most abundant species with low rates of occurrence (Goodbody-Gringley, pers. obs.). Furthermore, recruitment for most taxa is documented to decline dramatically below 50m, indicating that competitive exclusion has less influence on community structure at depth [47]. Thus, competition for space is likely not a limiting factor for population density on mesophotic reefs, as restricted light and nutrient availability reduces the abundance and diversity of competitive species allowing *M. cavernosa* to become well established [47, 48].

Species distribution and population structure are highly influenced by characteristics of the physical environment such as temperature and wave energy, which vary with depth and thus affect coral population dynamics on mesophotic reefs. Thermally induced coral bleaching is known to cause significant mortality on shallow-water reefs [49], however mesophotic corals appear to be well insulated from the effects of increased sea surface temperature (SST), which may be in part due to the lower degree of variability in SST experienced on mesophotic reefs compared with shallow reefs (Fig 7B) [8, 50, 51]. Although increased SST is a reliable indicator of increasing temperature at mesophotic depths >30m [52], Lesser and Slattery [51] report bleaching to be virtually absent on corals inhabiting mesophotic reefs. Reduced occurrences of bleaching events on mesophotic reefs is likely due to a lower maximum SST (Fig 7B) and solar isolation [53] at depth, despite the potential for cold-water stress to induce bleaching as reported elsewhere on Caribbean reefs [54–56]. Likewise, hydrodynamic disturbance and exposure to wave energy are major factors influencing community structure in shallow reef systems. Such disturbances are minimal on mesophotic reefs, however, as surface wave energy attenuates with depth [57, 58]. Corals inhabiting mesophotic reefs are, therefore, buffered from direct physical damage from rough hydrodynamic conditions, which may contribute to the long-term stability of these ecosystems, although episodic storm events may cause fragmentation of branching, foliose and columnar coral colonies at depth [9, 59]. Additionally, human-mediated stresses appear to be reduced on mesophotic reefs due primarily to increased distance from human populations and greater depths than nearby shallow reef systems [60]. Therefore, corals inhabiting mesophotic zones may be protected from biotic and abiotic impacts that typically occur on shallow-water coral reefs.

In fact, the results of this study indicate that colonies of *M. cavernosa* appear to form relatively stable populations on mesophotic reefs in Bermuda. Mean population size structure was bell-curved (Fig 4), and standard deviation, skewness, and kurtosis did not vary greatly by site (Table 2). These results suggest that the mesophotic zone, which extends around the perimeter of the Bermuda platform, creates a viable habitat able to support an established population of *M. cavernosa*.

Likewise, the mean size-frequency distribution of shallow reef populations was also bell curved, however, the overall size structure of shallow sites was shifted towards larger individuals with the smallest size classes underrepresented (Fig 4). Previous studies on coral population structures suggest that environmental deterioration may skew populations towards a greater proportion of larger individuals [13, 19]. While the results of the present study may indicate that the shallow reef environment in Bermuda is less stable than mesophotic regions, there was no statistical difference in skewness of the populations preventing any conclusive remarks as to the stability of shallow sites versus mesophotic sites.

These findings support previous survey work conducted with submersibles and ROV’s in other regions that show stable populations of scleractinian corals on mesophotic reefs, which have not undergone declines similar to those seen on their shallow water counterparts [6]. This apparent stability has led to the development of the “Deep Reef Refugia Hypothesis”, which
posits that coral populations at depths greater than 30m could serve as a source/sink for genetic diversity and future repopulation of shallow regions [9]. Several recent studies have undertaken comparisons of conspecifics at neighboring deep and shallow reefs, and show that while a slight degree of genetic discontinuity appears to be present at certain locations, other shallow/deep populations display evidence of genetic connectivity, supporting the possibility of repopulation of deteriorating shallow reefs by deep reef populations [61–63]. Understanding the degree of genetic connectivity among shallow and mesophotic corals will, therefore, ultimately indicate the ability of deep reefs to contribute to shallow reef resilience. Likewise, determining the health and stability of mesophotic coral populations through demographic analyses will suggest the viability of these reefs to serve as a source of propagules to maintain shallow water reefs and help guide future management and conservation strategies [25].

The results presented here represent a baseline assessment of coral population structure and reef condition on MCE’s in Bermuda. As the technology of mixed-gas closed circuit diving advances, it is anticipated that research on MCE’s will rapidly increase. Access to baseline data on community structure and reef condition will be imperative for future examinations of population demography, assessments of connectivity, projections of ecosystem change, and the overall resilience of global coral reef systems.

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Author Contributions

Conceived and designed the experiments: GGG ADC CM SG. Performed the experiments: GGG CM ADC. Analyzed the data: GGG CM. Contributed reagents/materials/analysis tools: GGG ADC. Wrote the paper: GGG CM ADC SG.

References

1. Jackson JBC, Donovan M, Cramer K, VV L. Status and trends of Caribbean coral reefs: 1970–2012. Gland, Switzerland: Global Coral Reef Monitoring Network, IUCN; 2014.
2. Pandolfi JM, Bradbury RH, Sala E, Hughes TP, Bjorndal KA, Cooke RG, et al. Global trajectories of the long-term decline of coral reef ecosystems. Science. 2003; 301:955–8. PMID: 12920296
3. Hughes TP, Rodrigues LJ, Bellwood DR, Ceccarelli D, Hoegh-Guldberg O, McCook L, et al. Phase shifts, herbivory, and the resilience of coral reefs to climate change. Curr Biol. 2007; 17:360–5. PMID: 17291763
4. Carpenter KE, Abrar M, Aeby G, Aronson RB, Banks S, Bruckner A, et al. One-third of reef-building corals face elevated extinction risk from climate change and local impacts. Science. 2008;560–3. doi: 10.1126/science.1159196 PMID: 18653892
5. Wilkinson CR. Coral reefs are facing widespread extinctions: can we prevent these through sustainable management practices? Proc 7th Int Coral Reef Symp. 2003; 1:11–21.
6. Bak RPM, Nieuwland G, Meesters EH. Coral reef crisis in deep and shallow reefs: 30 years of constancy and change in reefs of Curacao and Bonaire. Coral Reefs. 2005; 24:475–9.
7. Kahng SE, Copus JM, Wagner D. Recent advances in the ecology of mesophotic coral ecosystems (MCEs). J Exp Mar Biol Ecol. 2014; 7:71–81.
8. Lesser MP, Slattery M, Leichter JJ. Ecology of mesophotic coral reefs. J Exp Mar Biol Ecol. 2009; 375:1–8.
9. Bongaerts P, Ridgway T, Sampayo EM, Hoegh-Guldberg O. Assessing the deep reef refugia hypothesis: focus on Caribbean reefs. Coral Reefs. 2010; 29:309–27.
10. Hinderstein LM, Marr JCA, Martinez FA, Dowgiallo MJ, Puglise KA, Pyle RL, et al. Theme section on "Mesophotic Coral Ecosystems: Characterization, Ecology and Management". Coral Reefs. 2010; 29:247–51.
11. Clarke GL. Light penetration in the Western North Atlantic and its application to biological problems. Rapp. p.-v. réun.—Cons. int. explor. mer. 1936; 101:3–14.
12. Menza C, Kendall M, Hile S. The deeper we go the less we know. Rev Biol Trop. 2008; 56:11–24.
13. Bak RPM, Meesters EH. Coral population structure: the hidden information of colony size-frequency distributions. Mar Ecol Prog Ser. 1998; 162:301–6.
14. Hughes TP, Jackson JBC. Population dynamics and life histories of foliaceous corals. Ecol Monogr. 1985; 55:141–66.
15. Hughes TP, Jackson JBC. Do corals lie about their age? Some demographic consequences of partial mortality, fission and fusion. Science. 1980; 209:713–4. PMID: 17821194
16. Chadwick-Furman NE, Goffredo S, Loya Y. Growth and population dynamic model of the reef coral *Fungia granulosa* Klunzinger, 1879 at Eilat, northern Red Sea. J Exp Mar Biol Ecol. 2000; 249:199–218. PMID: 10841935
17. Meesters EH, Wesseling I, Bak RPM. Partial mortality in three species of reef-building corals and the relation with colony morphology. Bull Mar Sci. 1996; 58:838–52.
18. Meesters EH, Wesseling I, Bak RPM. Coral colony tissue damage in size species of reef-building corals: partial mortality in relation with depth and surface area. J Sea Res. 1997; 37:131–44.
19. Meesters EH, Hilterman M, Kardinaal E, Keetman M, de Vries M, Bak RPM. Colony size-frequency distributions of scleractinian coral populations: spatial and interspecific variation. Mar Ecol Prog Ser. 2001; 209:43–54.
20. Caroselli E, Zaccanti F, Mattioli G, Falini G, Levy O, Dubinsky Z, et al. Growth and demography of the solitary scleractinian coral *Leptosamma pruvoti* along a sea surface temperature gradient in the Mediterranean Sea. PLoS ONE. 2012; 7(6):e37848. doi:10.1371/journal.pone.0037848 PMID: 22675495
21. Fricke HW, Meischner D. Depth limits of Bermudan scleractinian corals: a submersible survey. Mar Biol. 1985; 88:175–87.
22. Szmant AM. Reproductive ecology of Caribbean coral reefs. Coral Reefs. 1986; 5:43–53.
23. Veron J. Corals of the World. Stafford-Smith M, editor. Townsville, Australia: Australian Institute of Marine Science; 2000. 490 p.
24. Reed JK. Deepest distribution of Atlantic hermatypic corals discovered in the Bahamas. Proc 5th Int Coral Reef Symp. 1985; 6:249–54.
25. Lesser MP, Slattery M, Stat M, Ojimi M, Gates RD, Grottoli A. Photoacclimation by the coral *Montastrea cavernosa* in the mesophotic zone: light, food and genetics. Ecology. 2010; 91:990–1003. PMID: 20462114
26. Davies PS. Respiration in some Atlantic reef corals in relation to vertical distribution and growth form. Biol Bull. 1980; 158:187–94.
27. Lasker HR. Phenotypic variation in the coral *Montastrea cavernosa* and its effects of colony energetics. Biol Bull. 1981; 160:292.
28. Bendel RB, Higgins SS, Teberg JE, Pyke DA. Comparison of skewness coefficient, coefficient of variation, and Gini coefficient as inequality measures within populations. Oecologia. 1989; 78:394–400.
29. Sokal RR, Rohlf FJ. Biometry. The principles and practice of statistics in biological research, 3rd edn. New York: WH Freeman & Company; 1995.
30. Sokal RR, Rohlf FJ. Biometry, 2nd edn. New York: WH Freeman and Company; 1981.
31. Shapiro S, Francia R. An appropriate analysis of variance test for normality. J Am Statist Assoc. 1972; 67:215–6.
32. van Veenbergen R, Done T. Coral communities and reef growth in the southern Great Barrier Reef. Coral Reefs. 1997; 16:103–15.
33. McClanahan T, Ateweberhan M, Omukoto J. Lont-term changes in coral colony size distributions on Kenyan reefs under different management regimes and across the 1998 bleaching event. Mar Biol. 2008; 153:755–68.
34. Anderson M, Gorley R, Clarke K. PERMANOVA+ for PRIMER: Guide to software and statistical methods. Plymouth, United Kingdom: PRIMER-E Ltd; 2008.
35. Fricke HW, Vareschi E, Schlichter D. Photoecology of the coral *Leptoseris fragilis* in the Red Sea twilight zone (an experimental study by submersible). Oecologia. 1987; 73:371–81.
36. Maragos JE, Jokiel PL. Reef corals of the Johnston Atoll: one of the world’s most isolated reefs. Coral Reefs. 1986; 4:141–50.
37. Kahng SE, Maragos JE. The deepest, zooxanthellate scleractinian corals in the world? Coral Reefs. 2006; 25:245.
38. Barnes D, Chalker B. Calcification and photosynthesis in reef-building corals and algae. In: Dubinsky Z, editor. Ecosystems of the world 25: coral reefs. Amsterdam: Elsevier; 1990. p. 109–31.
39. Falkowski PG, Jokiel PL, Kinsey RR. Irradiance and corals. In: Dubinsky Z, editor. Ecosystems of the world 25: coral reefs. Amsterdam: Elsevier; 1990. p. 89–107.
40. Anthony KRN, Hoegh-Guldberg O. Variation in coral photosynthesis, respiration, and growth characteristics in contrasting light microhabitats: an analogue to plants in forest gaps and understoreys. Funct Ecol. 2003; 17:246–59.
41. Grigg RW. Depth limit for reef building corals in the Au’au Channel, S.E. Hawaii. Coral Reefs. 2006; 25:77–84.
42. Crabbe MJC. Scleractinian coral population size structures and growth rates indicate coral resilience on the fringing reefs of North Jamaica. Mar Environ Res. 2009; 67:189–98. doi:10.1016/j.marenvres.2009.01.003 PMID: 19269026
43. Bongaerts P, Frade PR, Hay KB, Englebert N, Latijnhouwers KRW, Bak RPM, et al. Deep down on a Caribbean reef: lower mesophotic depths harbor a specialized coral-endosymbiont community. Scientific Reports. 2015; 5(7652).
44. Bak RPM. The growth of coral colonies and the importance of crustose coralline algae and burrowing sponges in relation to carbonate accumulation. Neth J Sea Res. 1976; 10:285–337.
45. Van Moorsel GWM. Disturbance and growth of juvenile corals (Agaricia humilis and Agaricia agaricites, Scleractinia) in natural habitats on the reef of Curacao. Mar Ecol Prog Ser. 1985; 24:99–112.
46. Venn AA, Weber FK, Loram JE, Jones RJ. Deep zooxanthellate corals at the high latitude Bermuda Seamount. Coral Reefs. 2009; 28:135.
47. Avery WE, Liddell WD. Sessile community recruitment patterns on shallow- and deep-reef hard substrata. Proc 8th Int Coral Reef Sym. 1997; 2:1179–84.
48. Liddell WD, Avery WE. Temporal change in hard substrate communities 10–250 m, the Bahamas. Proc 10th Int Coral Reef Symp. 2000; 1:437–42.
49. Hoegh-Guldberg O, Mumby PJ, Hooten AJ, Steneck R, Greenfield P, Gomez E, et al. Coral reefs under rapid climate change and ocean acidification. Science. 2007; 318:1737–42. PMID: 18079392
50. Bongaerts P, Bridge TCL, Kline D, Muir P, Wallace C, Beaman R, et al. Mesophotic coral ecosystems on the walls of Coral Sea atolls. Coral Reefs. 2011; 30:335.
51. Lesser MP, Slattery M. Phase shift to algal dominated communities at mesophotic depths associated with lionfish (Pterois volitans) invasion on a Bahamian coral reef. Biol Invasions. 2011; 13:1855–68.
52. Donner SD, Skirving WJ, Little CM, Oppenheimer M, Hoegh-Guldberg O. Global assessment of coral bleaching and required rates of adaptation under climate change. Global Change Biol. 2005; 11:2251–65.
53. Loya Y, Sakai K, Yamazato K, Nakano Y, Sambali H, Woesik Rv. Coral bleaching: the winners and the losers. Ecol Lett. 2001; 4:122–31.
54. Lirman D, Schopmeyer S, Manzella D, Gramer LJ, Precht WF, Muller-Karger F, et al. Severe 2010 cold-water event caused unprecedented mortality to corals of the Florida reef tract and reversed previous survivorship patterns. PLoS ONE. 2011; 6(8):e23047. doi: 10.1371/journal.pone.0023047 PMID: 21853066
55. Menza C, Kendall M, Rogers C, Miller J. A deep reef in deep trouble. Cont Shelf Res. 2007; 27:2224–30.
56. Kobluk DR, Lysenko MA. “Ring” bleaching in southern Caribbean Agaricia agaricites during a rapid water cooling. Bull Mar Sci. 1994; 54:142–50.
57. Huston MA. Patterns of species-diversity on Coral Reefs. Ann Rev Ecol Syst. 1985; 16:149–77.
58. Kahng SE, Garcia-Sais JR, Spalding HL, Brokovich E, Wagner D, Weil E, et al. Community ecology of mesophotic coral reef ecosystems. Coral Reefs. 2010; 29:255–75.
59. White KN, Ohara T, Fuji T, Kawamura I, Mizuyama M, Montenegro J, et al. Typhoon damage on a shallow mesophotic reef in Okinawa, Japan. PeerJ. 2013; 1:e151. doi: 10.7717/peerj.151 PMID: 24032094
60. Riegl B, Piller WE. Possible refugia for reefs in times of environmental stress. Int J Earth Sci. 2003; 92:520–31.
61. Serrano X, Baums IB, O'Reilly K, Smith TB, Jones RJ, Shearer TL, et al. Geographic differences in vertical connectivity in the Caribbean coral *Montastraea cavernosa* despite high levels of horizontal connectivity at shallow depths. Mol Ecol. 2014; 23:4226–40. doi: 10.1111/mec.12861 PMID: 25039722

62. Bongaerts P, Riginos C, Ridgway T, Sampayo EM, Van Oppen MJH, Englebert N, et al. Genetic divergence across habitats in the widespread coral *Seriatopora histrix* and its associated *Symbiodinium*. PLoS ONE. 2010; 5(5):e10871. doi: 10.1371/journal.pone.0010871 PMID: 20523735

63. Brazeau DA, Lesser MP, Slattery M. Genetic structure in the coral *Montastraea cavernosa*: assessing genetic differentiation among and within mesophotic reefs. PLoS ONE. 2013; 8(5):e65845. doi: 10.1371/journal.pone.0065845 PMID: 23734263