Performance and environmental accounting of nutrient cycling models to estimate nitrogen emissions in agriculture and their sensitivity in life cycle assessment

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Abstract
Purpose  Several models are available in the literature to estimate agricultural emissions. From life cycle assessment (LCA) perspective, there is no standardized procedure for estimating emissions of nitrogen or other nutrients. This article aims to compare four agricultural models (PEF, SALCA, Daisy and Animo) with different complexity levels and test their suitability and sensitivity in LCA.

Methods  Required input data, obtained outputs, and main characteristics of the models are presented. Then, the performance of the models was evaluated according to their potential feasibility to be used in estimating nitrogen emissions in LCA using an adapted version of the criteria proposed by the United Nations Framework Convention on Climate Change (UNFCCC), and other relevant studies, to judge their suitability in LCA. Finally, nitrogen emissions from a case study of irrigated maize in Spain were estimated using the selected models and were tested in a full LCA to characterize the impacts.

Results and discussion  According to the set of criteria, the models scored, from best to worst: Daisy (77%), SALCA (74%), Animo (72%) and PEF (70%), being Daisy the most suitable model to LCA framework. Regarding the case study, the estimated emissions agreed to literature data for the irrigated corn crop in Spain and the Mediterranean, except N2O emissions. The impact characterization showed differences of up to 56% for the most relevant impact categories when considering nitrogen emissions. Additionally, an overview of the models used to estimate nitrogen emissions in LCA studies showed that many models have been used, but not always in a suitable or justified manner.

Conclusions  Although mechanistic models are more laborious, mainly due to the amount of input data required, this study shows that Daisy could be a suitable model to estimate emissions when fertilizer application is relevant for the environmental study. In addition, and due to LCA urgently needing a solid methodology to estimate nitrogen emissions, mechanistic models such as Daisy could be used to estimate default values for different archetype scenarios.

Keywords  IPCC TIERs · UNFCCC · Nitrate leaching · Ammonia volatilization · Nitrous oxide · PEF · Daisy · Animo · SALCA IPCC TIERs

1 Introduction

Appropriate resource management in agricultural systems is the responsibility and a challenge of the agronomic sector and environmental policies, especially to match growing demand and crop production (Wuepper et al. 2020). The objective of agricultural production is to provide safe and good quality food in such a way to minimize adverse impacts on the environment. To sustain food production, around 75% of the reactive nitrogen added to agroecosystems is created by human activities, and the excess of nutrients is a severe problem and threatens the environmental
balance (Rockström et al. 2009). In particular, nitrogen (N) emissions to air, soil and water may have several adverse effects. For instance, climate change is affected by nitrous oxide (N_2O) emissions, and nitrogen oxides (NO_x) form acid when interacting with water, oxygen and other chemicals, contributing to acidification (Frischknecht and Jolliet 2016). In the same way, marine eutrophication is the consequence of nitrate (NO_3^-) emission exposure to aquatic systems (Wolf 2010) and pollution of groundwater due to NO_3^- leaching may cause a decrease in freshwater resource quality and hence affect human health (Ward et al. 2018).

For assessing impacts from agriculture, the life cycle assessment (LCA) is a broadly accepted and used methodology (Notarnicola et al. 2017; Nitschelm et al. 2018). Agricultural systems LCA can use LCA to calculate the environmental costs on goods and services by quantifying all emissions and resource consumption. However, to use LCA, there is a need to estimate the sources of nutrients (e.g. nitrogen, phosphorus) responsible for the most significant impacts on the environment (Groenendijk et al. 2005).

According to input data needs and the degree of complexity, the IPCC (2006) classifies in three different tiers, the methodological approaches for estimating nutrient emissions. Models that are considered Tier 1 use the default emission factors (EF) provided, for instance, by IPCC. Tier 2 models are very similar to Tier 1, but EFs and other parameters applied are country-specific. Tier 3 models are the most detailed; therefore, it can estimate the emissions with greater certainty than Tiers 1 and 2.

While there is no standardized methodology or models to estimate nutrient emissions in LCA, many methodologies have been used. Brentrup et al. (2000) proposed Tier 1 and Tier 2 models to estimate the most important nitrogen emissions (NH_3, N_2O, NO_3^-) related to agricultural production in LCA. Tier 2 models, for instance, SALCA (Nemecek et al. 2016) and AGRYBALYSE (Koch and Salou 2015), and Tier 3 models, such as DAYCENT in Kim and Dale (2005), DNDC in Goglio et al. (2014) and STICS in Plaza-Bonilla et al. (2018), have also been used to estimate nitrogen emissions in LCA.

The guideline “Nutrient flows and associated environmental impacts in livestock supply chains” (FAO 2018) provides recommendations for building inventories in life cycle assessment (LCA) regarding the level of specificity of the study. Tier 1 is recommended for a screening analysis that allows the practitioner to overview the hotspots in the studied system. Tier 2 is recommended for supply chain and regional assessments, and Tier 3 should be applied to the product system. However, since those are recommendations, LCA practitioners are not forced to choose one model or other, but, for example, as Perrin et al. (2014) claimed, models used to estimate emissions can sometimes be used in inappropriate domains they were created.

In this sense, two Tier 3 dynamic models Animo and Daisy, the Tier 2 LCA emission method SALCA (Nemecek et al. 2015), and the Product Environmental Footprint (PEF) (EC-PEFCR 2018) were applied to estimate nitrogen emissions from agriculture in LCA. The two dynamic Tier 3 models, Animo (Rijtema and Kroes 1991) and Daisy (Hansen 2000), have been used to estimate the nitrogen emissions to soil, air and water under the scope of the European Union’s Horizon 2020 Project Nutri2Cycle (Grant agreement No 773682, https://www.nutri2cycle.eu/). The different models (for terminology consistency, all approaches will be referred to as models) are compared and discussed, considering their requirements and main characteristics. The specific aims of this study can be divided into the following:

1. Provide an overview of the selected models to understand their main characteristics and application in agricultural systems;
2. Compare PEF, SALCA, Daisy and Animo under the adapted criteria from the United Nations Framework Convention on Climate Change (UNFCCC) and other relevant studies to judge their suitability in LCA framework;
3. Perform a quantitative comparison using an irrigated maize production case study in Spain. Additionally, impacts were characterized considering the different emissions estimated;
4. Discuss how nitrogen emissions have been estimated in LCA agricultural studies found in literature and suggest how nutrient emission models should be used in LCA.

2 Methods

The assessment of the different emission accounting models included several steps:

1. Contextual background of the models (Sect. 2.1);
2. Introduction to the N cycle and its consideration and adaptation in the models (Sect. 2.2);
3. Description of criteria and subcriteria for the models’ evaluation (Sect. 2.3);
4. Description of the case study performed (Sect. 2.4).

2.1 Contextual background of the models

In this section, an overview of the models is provided, also their application in agricultural systems.

The European Commission’s and the Joint Research Center (JRC) developed the PEF model. The Swiss Confederation center for agricultural research (Agroscope) developed and recommended methods that established SALCA. The Agrohydrology group at the University of
Copenhagen developed the mechanistic simulation model of agricultural field model, Daisy, and Wageningen University and Research is the institution behind Animo model.

Regarding spatial scale, Daisy and Animo present the most detailed scale, site-specific nutrient emissions. SALCA appears to be the most limited in reproducing emission estimates, due to its focus on crops and farms in Europe or in temperate climate zone. PEF does not cover spatial scale.

SALCA, Daisy and Animo provide default crop parameters in the models’ library. These default values are crucial for LCA practitioners who wish to use dynamic models to estimate emissions. Still, they do not have sufficiently detailed information to create a new crop dataset. The common crops simulated in all models are maize, potatoes, grassland and wheat.

One way to judge the accuracy and precision of a model is through validation of its parameters. Those parameters may come from field observations, model calibration, or user expertise (Hansen et al. 2012). Model calibration in Animo and Daisy can use yield. A simplified validation of the results can be made based on literature data from other studies, on similar conditions. PEF, SALCA, Daisy and Animo have already been calibrated and validated under different climatic types defined by Koppen-Geiger (Table 1).

PEF and SALCA are considered user-friendly models, due to its simplicity (PEF), adaptation to spreadsheets and use of parameters from literature (SALCA). Although Animo and Daisy cannot be considered as user-friendly models, due to the programming and the amount of input data required for the models, spreadsheet files or text editors are used to read their outputs.

All models provide a compiled bibliography (i.e. user guide, references, tutorial), which is especially helpful for non-experts or the beginners in the models. Moreover, Daisy offers strategies to deal with the lack of data, guiding users to minimize the effect of the assumptions on results and providing user support to help understand the model and the simulations performed. Strategies for unavailable data and user support for SALCA and, especially, for Animo would be useful for the practitioners.

Regarding the suitability of the models in LCA, SALCA and PEF were explicitly developed for LCA studies. Daisy and Animo are compatible with the LCA methodology since they provide the necessary emissions. Daisy was used to estimating emissions in LCA for garden waste management options (ten Hoeve et al. 2019), to quantify greenhouse gas emissions (Jensen et al. 2017), and estimate emissions in Danish cereal cropping systems (Kløverpris et al. 2016). SALCA, initially developed for Switzerland, has been extended to other countries with a temperate climate and has been used in several European projects that include LCA in its scope. PEF has already been used to assess the environmental performance of different agricultural products such as wines, pasta and dairy products. Animo has not yet been used in LCA.

Uncertainty and sensitivity analyses are fundamental in LCA studies because it can estimate emission ranges for results and can develop scenarios appropriately. SALCA is the only model that does not consider the sensitivity and uncertainty of their parameters. The uncertainty and sensitivity of IPCC emission factors are considered for PEF. The uncertainty in Daisy was evaluated for the input parameters, obtaining a range between 5 to 95% comparing the measured monthly soil water content and the estimates from the model (Salazar et al. 2013). Jabloun et al. (2016) analysed the sensitivity of the outputs showing that the weather conditions substantially influence the Daisy’s outputs. Kroes and Roelsma (2007) evaluated the uncertainty related to the hydraulic parameters (measured and estimated) in Animo and concluded that there is a little influence (< 3% changes) on nitrate leaching. Hendriks et al. (1999) focused on solute transport adaptations in Animo, where demonstrated high sensitivity to oxygen diffusion parameters and can influence nitrogen processes such as mineralization, nitrification and denitrification.

| Climate                          | PEF | SALCA | Daisy | Animo            |
|----------------------------------|-----|-------|-------|------------------|
| Tropical/megathermal            | Y*  | N     | N     | Pinto (2016)     |
| Dry (desert and semiarid)       | Y*  | N     | Manevski et al. (2016) | Farmaha (2014) |
| Temperate/mesothermal           | Y*  | N     | Mueller et al. (1997) | Rijtema and Kroes (1991) |
| Continental/microthermal        | Y*  | N     | Pohanková et al. (2015) | Marinov et al. (2005) |
| Polar                           | Y*  | N     | N     | N                |
| Extreme weather conditions      | Y*  | N     | N     | Hendriks and Akker (2017) |

*The PEF was created to be used worldwide, and there is no restriction for application in different climate conditions.
2.2 Introduction to the N cycle and its adaptation in the models

In this section, the models’ consideration of processes in N cycle is explained. In addition, the critical N emissions for the Life Cycle Inventory (LCI), namely, nitrification (N2O and N2O) and NOx leaching (NOx), denitrification (N2 and N2O) and volatilization (NH3) are detailed (Fig. 1) (Table 2).

Nitrate (NO3-) leaching in agriculture can occur when excess nitrate fertilizer is applied and lost due to rain or irrigation, among other soil and crop properties, and through aerobic microbially driven nitrification of ammonium ions. NO3- leaching is estimated in PEF, using the EF 0.44 kg NO3-/kg N and the amount of fertilizer applied. In SALCA, this estimate is made using a balance between inputs (fertilization and irrigation) and outputs (plant uptake and background nitrogen emissions) using simplified equations. The process is more complex in Daisy and Animo, where nitrate inputs come from atmospheric deposition, fertilizers and soil solution. They apply a water-balance model using Darcy’s law (Cannavo et al. 2008).

Ammonia volatilization (NH3) occurs typically when the nitrogen is in the form of urea, which can come mainly from animal manure or urea fertilizers. All models estimate NH3 volatilization in a similar yet limited way, applying EF or volatilized fertilizer fractions. In PEF, different EFs (kg N/kg N applied) are used, for instance, 0.15 for urea and 0.1 for ammonium nitrate. In SALCA, NH3 emissions depend on the type and quantity of fertilizer, N content of the fertilizer, pH and the air saturation deficit. In Animo and Daisy, volatilization is not a function of climate conditions or incorporation depth. Thus, the user must enter a value for a fraction of NH4+ that evaporates after applying the fertilizer. It is important to highlight that only Animo takes into account the fertilizer application practices (e.g. broad sprayer, hose, injection), illustrating a limitation in the other models since many studies have found that practices can influence NH3 volatilization (Bittman et al. 2014; Søgaard et al. 2002; Brentrup et al. 2000; with an example of its site-specific application and use in Montemayor et al. 2019).

Nitrous oxide (N2O) emitted by soils can be produced by denitrification in anoxic conditions or by nitrification in the presence of O2, being an intermediate emission of incomplete nitrification and denitrification reactions. In PEF, N2O is estimated using the IPCC (2006) modified EF of 0.022 (kg N2O/kg N applied). SALCA considers direct (from nitrogen oxide (NO-N)) and indirect (from NH3 and NOx) N2O emissions, using the EF of 0.01 (kg N2O-N/kg N applied) for that. N2O is estimated by Michaelis-Menten kinetics in Daisy, depending on the availability of NH4+ and general heterotrophic respiration. In Animo, N2O is estimated by an empirical equation that depends, among other parameters, on the concentration of NH4+, the water content in the layer, temperature and pH.

![Fig. 1 Nitrogen cycle and main processes (adapted from Abrahamsen and Hansen 2000)](image1)

| Parameter                  | PEF | SALCA | Daisy | Animo |
|----------------------------|-----|-------|-------|-------|
| Nitrogen fixation          |     | x     | x     | x     |
| Decomposition              | x   | x     | x     | x     |
| Immobilization/mineralization | x | x | x | x |
| Nitrification              | x   | x     | x     | x     |
| Atmospheric deposition      |     | x     | x     | x     |
| Ammonium leaching          | x   | x     | x     | x     |
| Ammonium adsorption/desorption |   | x | x | x |
| Plant uptake               | x   | x     | x     | x     |
| Nitrate leaching           | x   | x     | x     | x     |
| Denitrification            | x   | x     | x     | x     |
| Volatilization             |     | x     | x     | x     |
Denitrification is the process by which NO$_3^-$ is reduced to N$_2$ in a total reduction or NO$_2$ and N$_2$O in a partial reduction. In PEF, total denitrification producing N$_2$ is assumed using the EF 0.09 kg N$_2$/kg N applied. Denitrification is not included in SALCA (Nemecek et al. 2016). In Daisy, denitrification is affected by temperature and water pressure and depends on a maximum fraction of converted nitrate, among other factors. Denitrification in Animo is considered a partial or complete reduction of available nitrate, depending on the respiration of organic matter, biodegradable organic matter, soil layer thickness and nitrate concentration. A denitrification rate is also required for limited nitrate conditions in Animo. For NO$_x$ emissions, SALCA uses IPCC (2006) EF, 0.012 (kg NO$_x$-N/kg N applied), while PEF, Daisy and Animo do not estimate NO$_x$ emissions.

Other parameters not detailed in this section can directly and or indirectly affect the N emissions estimations. For instance, in mineralization, nutrients released as soluble inorganic bioavailable forms, and the roots’ nitrogen uptake establish a balance between the crop’s demand and the supply by the soil. Equations available are in the Supplementary Material.

### 2.3 Description of the models and applied comparison metrics

A set of different criteria and sub-criteria, based on UNFCCC (2004), Vidal-Legaz et al. (2016) and International Life Cycle Data (ILCD) (Wolf et al. 2010), were proposed to score and rank the models according to their user-friendliness and applicability for use as agricultural emission models in LCA studies. The criteria included are ‘completeness of the model scope’, ‘environmental relevance’, ‘scientific robustness’, ‘availability, documentation, transparency and reproducibility’, ‘applicability and flexibility’ and ‘stakeholder acceptance’ (Table 3). The possible scores were 1 (poor), 3 (good) and 5 (excellent).

### 2.4 Case study: maize crop in Spain

A case study was used to compare the estimates calculated using the models. A scenario of irrigated maize (2013–2017) in Mediterranean climate using calcium ammonium nitrate (CAN) as fertilizer was used (Table 4).

The minimum parameters required to estimate N emissions in the models are shown in Table 5 and Supplementary Material 2. Concerning Daisy and Animo’s set up, a calibration was provided to align the models’ outputs with real field measurements using yields from maize crop rotations (2013–2017). Default values for parameters in Daisy and Animo were taken from the models’ library.

Note that although PEF seeks to standardize emissions for certain agricultural products, the low amount of input data required to estimate N emissions can result in lower accuracy and representativeness. It is important to highlight also that the pilot phase of PEF did not include cultivation in the foreground system, but it is under review for future assessments.

In Daisy, to reduce the effect of extreme weather conditions, a simulation was done for a 100-year simulation, applying randomized weather-crop combinations. In Animo, a 5-year simulation was performed to initialize an adjusted soil organic matter pool (SOM) for better estimates in the model.

An automatic irrigation (30 mm/h in case the water pressure in the soil falls below − 600 cm in the top 30 cm soil from May to September) had to be used in Daisy due to the impossibility to perform irrigation on specific days, as used in Animo. The nitrogen supplied by irrigation in SALCA was calculated multiplying the concentration of N in the water irrigation and total irrigation applied. Irrigation in SALCA was taken into account, adding it to monthly precipitation, in order to select a coefficient for soil leaching. For PEF, neither the N in irrigated water nor irrigation are considered.

Regarding the N estimates provided by Daisy and Animo, NO$_3^-$ leaching was calculated for the 100 cm, depth of the root zone. N$_2$O (nitrification and denitrification) and NH$_3$ were estimated for the total soil profile. NO$_3^-$ leaching in SALCA was estimated for 90 cm of depth.

In the present work, the nitrogen balance in the field from the results obtained with the models includes as inputs: the mineral and organic fractions of fertilizers, atmospheric deposition, N in the irrigation water and fixation of atmospheric N by legumes. As sources of N production are losses to groundwater and surface water (via leaching and nitrate runoff), emissions to the atmosphere via ammonia volatilization, nitrification (N$_2$O and denitrification) and N absorption by crops and harvested N. The stock of N (N inputs minus N outputs) in the soil is a positive value (increasing) that indicates the input N is greater than the output, contributing to the increase in the stock of N. Otherwise, if the change in the stock of N is a negative (decreasing) value suggests mineralization of organic N from the soil. Therefore, the crop is taking nitrogen out of the soil. The strategy adopted for the N balance is the same used in the Daisy and Animo models.

It is essential to highlight that emissions estimated in dynamic models day by day use precisely climate condition for the management operation performed, but much more detailed information is required, which can be an obstacle for LCA practitioners. The simulations made in Animo and Daisy were carried out in the most similar way possible, but, due to models’ internal parameters, differences were found in the results provided.

The estimated emissions were inventoried in SimaPro software v. 8.5 (Pré Consultants 2017) using a scenario provided by Montemayor et al. (2019). The impacts were characterized using the ILCD 2011 midpoint method to verify how variations in emissions estimations influence LCA impact results.
| Criteria                                                                 | Description and scoring                                                                 |
|--------------------------------------------------------------------------|------------------------------------------------------------------------------------------|
| Completeness of the model scope                                           |                                                                                          |
| Geographic coverage                                                      | 1 = Local; 3 = Regional; 5 = Global                                                        |
| Environmental relevance                                                  |                                                                                          |
| Spatial-temporal resolution Temporal resolution of the input              | 1 = Annual; 3 = Seasonal; 5 = Monthly or higher resolution                                |
| Spatial resolution of the input                                           | 1 = Global; 3 = Regional/National; 5 = Municipality/farmer scale                          |
| Scientific robustness                                                    |                                                                                          |
| Transparency                                                              | 1 = No clear modelling explanation, not easily understood; 3 = Processes are clearly modelled but not easily understood; 5 = Processes are clearly modelled and easily understood |
| Input data set/data requirements                                          | 1 = Extensive and detailed input parameters needed; 3 = Application of a questionnaire in a farm, a simple dataset for meteorological and soil physical parameters; 5 = Small and basic parameter input, data obtained global databases or literature |
| Emission model peer-review and (peers) acceptance                        | 1 = No (unpublished report); 3 = Partially (book or authoritative body report with some review process, or partial publication in a journal, including all parts of the model); 5 = Yes (full peer-reviewed journal for all aspects of the model) |
| The model reflects up-to-date knowledge for the cause-effect chain        | 1 = not up-to-date; 3 = partially up-to-date; 5 = yes (state-of-the-art)                   |
| Tests of the emissions already conducted                                 | 1 = No; 3 = Tested for relevant products/scale and conditions but showing important limitations; 5 = Tested for relevant products/scale, different conditions, peer-reviewed and showing not relevant model limitations |
| Uncertainty and sensitivity analysis                                      | 1 = No; 3 = Yes, but just for the outputs; 5 = Yes, including inputs and outputs           |
| Availability, documentation, transparency and reproducibility            |                                                                                          |
| Accessibility of the emission model                                      | 1 = No free access/availability; 3 = Available under conditions/on request; 5 = Free access/Internet download |
| Accessibility of the model documentation                                 | 1 = Not accessible; 3 = Accessible with limitations (e.g. fee due, not available in the English language); 5 = Totally accessible |
| Accessibility of the input data                                          | 1 = High limitations (many input data not available in global databases, also data not related to common in LCA); 3 = Low limitations (some data too specific and not available in regional database); 5 = Totally accessible, all data are relatively easy to obtain |
| Modelling assumptions and value choices                                  | 1 = Not described; 3 = Unclear/partial description; 5 = Comprehensive description         |
| Completeness of the emission model documentation                          | 1 = Very incomplete or no documentation; 3 = Partially comprehensive documentation; 5 = Fully comprehensive documentation |
| Applicability and flexibility                                            |                                                                                          |
| Compatibility with LCA methodology                                       | 1 = Not compatible; 3 = Not developed for LCIA but it fits the scope; 5 = Developed for LCA and tested |
| Usability of models for LCA practitioners                                 | 1 = Not used; 3 = Already used but in few situations; 5 = Already used in several studies |
| Related to IPCC TIER concept                                             | 1 = Tier; 3 = Tier 2; 5 = Tier 3                                                          |
| Management operations consideration                                      | 1 = No; 3 = Partially includes; 5 = Totally includes                                      |
3 Results

3.1 Comparison under the criteria and sub-criteria proposed by UNFCCC and other authors for adequacy in the LCA studies

The model with the best total score was Daisy with 91 (79% of the maximum total score), followed by SALCA and Animo with 85 (74%) and PEF with 77 (67%). The percentage achieved for each model in the selected criteria is shown in Fig. 2 and Table 6. Detail scored will be explained in this section, and further elaborated in Supplementary Material.

3.1.1 ‘Completeness of the model scope’ and ‘Environmental relevance’

In ‘Geographic coverage’ sub-criteria, PEF, Daisy and Animo scored 5 due to their worldwide applicability. Daisy and Animo require a model calibration, and there is no spatial restriction for PEF due its simplicity. SALCA scored 3 because it was developed to estimate emissions in Europe or temperate climate in the Northern Hemisphere.

3.1.2 ‘Scientific robustness’

SALCA, Daisy and Animo models scored 5 for ‘Transparency’ because the processes to estimate nitrogen emissions were clearly modelled and well-explained, PEF scored 1 because the emission fractions used are not adapted to different climate conditions and the system processes for cultivation are non-transparent. It was assumed, also taking into account Nemecek et al. (2016), that the less the ‘Input dataset/data requirements’, the better the model, as it requires less effort and time from the LCA practitioner. Thus, PEF scored 1, SALCA 3 and Daisy and Animo 1 since the last two need more input data (Table 2).

Regarding ‘Emission model peer review and (peers) acceptance’, Daisy and Animo scored 5 because they are peer-reviewed. SALCA and PEF are provided as guidelines, receiving a score of 3.

Daisy and Animo explain the entire nitrogen cycle and the interconnections within the cycle, receiving 5 in ‘The model reflects up-to-date knowledge for the cause-effect chain’. SALCA and PEF failed to receive the maximum...
Daisy and Animo had ‘Tests of the emissions already conducted’, scoring 5. Although SALCA (3) has been used in LCA studies, it is not well-validated and has restricted use. The climate data in PEF (3) is not representative, so inconsistencies can be found compared with field measurements. PEF, Daisy and Animo scored 3 in ‘Uncertainty and sensitivity analyses’, and SALCA scored 1 because there is no information about those analyses for the model.

Table 5 Minimum parameters required to estimate nitrogen emissions using PEF, SALCA, Daisy and Animo

|                | PEF  | SALCA | Daisy | Animo |
|----------------|------|-------|-------|-------|
| **Weather data** | - None | - Average monthly precipitation | - Main characteristics of the weather station | - Depth of the horizons, of max rooting, groundwater and existence of drainage |
|                |      |       |       | For each soil horizon (A, B, C…) |
|                |      |       |       | - Clay (%), Silt (%) and Sand (%) |
|                |      |       |       | - Humus (%) |
|                |      |       |       | - C:N |
|                |      |       |       | - Bulk density |
|                |      |       |       | - Mualem van Genuchten model: |
|                |      |       |       | - α and n (shape parameters) |
|                |      |       |       | - Ksat (saturated hydraulic conductivity cm/d) |
| **Soil characteristics** | - None | - pH | - Slope | - Depth of the horizons, of max rooting, groundwater and existence of drainage |
|                |      |       | - N in soil | For each soil horizon (A, B, C…) |
|                |      |       | - Coefficient related to rain washing | - Clay (%), Silt (%) and Sand (%) |
|                |      |       | - Leaching coefficient as a function of the slope | - Humus (%) |
|                |      |       |       | - C:N |
|                |      |       |       | - Bulk density |
| **Fertilizer** | - Amount | - Type and amount | - Type and amount (for NH₃ emissions) | - α and n (shape parameters) |
|                | - Type | - N availability (organic fertilizers) | - Total C fraction (%) | - Ksat (saturated hydraulic conductivity cm/d) |
|                |       |       | - Total N fraction (%) | - NH₄⁺-N fraction (%) |
|                |       |       | - NH₄⁺ volatilization (emission fraction) | - NH₄⁺-N fraction (%) |
|               |       |       |       | - NH₄⁺ volatilization (emission fraction) |
| **Crops and field management activities** | - None | - N uptake (fraction) | - Type of crop | - Dry matter (ton DM/ha) |
|                |       |       | - N content in the water irrigation | - Date of ploughing, fertilization, sowing, irrigation and harvesting |
|                |       |       |       | - Information about storage organ (leaf, stem, stub) |
| **Average yields (annual)** | - None | - None | - Dry matter (ton DM/ha) | - Yield (ton/ha) |
|                |       |       |       | - N content (kg N/ha) |

Fig. 2 Comparison of PEF, SALCA, Daisy and Animo under adapted methodology proposed by UNFCCC (2004) (ADTR availability, documentation, transparency and reproducibility, CMS completeness of the model scope, ER environmental relevance, AF applicability and flexibility, SR scientific robustness)
Table 6  Detailed scores regarding the qualitative assessment for comparing PEF, SALCA, Daisy and Animo models

| Criteria and subcriteria                                      | Models |
|---------------------------------------------------------------|--------|
| Completeness of the model scope                              |        |
| Geographic coverage                                          | 5      |
| Environmental relevance                                      |        |
| Spatial-temporal resolution                                  | 1      |
| Temporal resolution of the input data                        | 5      |
| Spatial resolution of the input data                         | 5      |
| Scientific robustness                                        |        |
| Transparency                                                 | 1      |
| Input data set/data requirements                             | 5      |
| Emission model peer-review and (peers) acceptance            | 3      |
| The model reflects up-to-date knowledge for the cause-effect chain | 3  |
| Tests of the emissions already conducted                     | 3      |
| Uncertainty analysis                                         | 3      |
| Availability, documentation, transparency and reproducibility |        |
| Accessibility of the emission model                          | 5      |
| Accessibility of the characterization model documentation     | 5      |
| Accessibility of the input data                              | 5      |
| Modeling assumptions and value choices                       | 3      |
| Completeness of the emission model documentation             | 5      |
| Applicability and flexibility                                |        |
| Compatibility with LCA methodology                          | 5      |
| Usability of models for LCA practitioners                     | 5      |
| Related to IPCC TIER concept                                 | 1      |
| Management operations                                        | 1      |
| Flexibility                                                  | 1      |
| Model and model results                                      | 5      |
| Authoritative body                                           | 5      |
| Academic authority                                           | 1      |
| Neutrality across industries, products, or processes         | 5      |

3.1.3 ‘Availability, documentation, transparency and reproducibility’

PEF, SALCA and Daisy scored 5 because the models provide an easy ‘Accessibility of the emission model’. PEF and SALCA provide documentation, and Daisy is an executable program, run in a text editor, that can be downloaded from of the University of Copenhagen website and uses its own programming language. Animo, though also an executable program run in a text editor, scored 3 because a request for access to the model is necessary.

SALCA scored 3 in ‘Accessibility of the characterization model documentation’ because the model is only available in German, which may represent a language barrier for many LCA practitioners. Daisy, Animo and PEF scored 5 because they provide useful documentation for a complete understanding of the models.

PEF scored 5 in ‘Accessibility of the input data’, because the amount and type of fertilizer are the only input data required. SALCA scored 3 as it is easy to obtain input data considering the inventory already created for the LCA study. Daisy and Animo scored 1 because some specific values may be more challenging to obtain, for instance, soil horizon characteristics, data for the groundwater or specific data related to the crop (e.g. leaves and roots).

Daisy scored 5 because it provides a document for ‘Modelling assumptions and value choices’, to help the user in cases with lack of data. SALCA and Animo scored 3 because assumptions are outlined in their reference documents. PEF scored 3 because it is not clear how the assumptions are made in the model, possibly due its simplicity. PEF, Daisy and Animo scored 5 in ‘Completeness of the emission model documentation’ because all the information required is described in the manuals. SALCA scored 3 since the manual was written for specific spatial conditions.
3.1.4 ‘Applicability and Flexibility’

PEF and SALCA obtained the maximum score in ‘Compatibility with LCA methodology’ and ‘Usability of models for LCA practitioners’ because they were created to estimate emissions in LCA studies. Daisy and Animo scored 3 in the former criteria because they were not developed for LCA, but fall within the scope. Daisy scored 3 and Animo 1 in the latter subcriterion because Daisy has already been used to estimate emissions in LCA, but in Animo they did not implement that aspect.

For the subcriterion ‘Related to IPCC Tier concept’, it was assumed that the model that best includes the dynamics on the environment (Tier 3) is the best model for LCA. Thus, PEF scored 1, SALCA 3, Daisy and Animo 5.

PEF scored 1 in ‘Management Operations’ because they are not considered in the model. SALCA scored 3 because some (e.g. irrigation) are relevant for the model. Daisy and Animo scored 5, since management operations are crucial for the models’ performance.

PEF scored 1 in ‘Flexibility’ because the model applies EF as default methodology. However, in the guideline (EC-PEFCR 2018) it is said that other nitrogen field model can be used under certain conditions. SALCA scored 5, because changes and assumptions in the model are easy to carry out since the model is based on equations. Daisy and Animo scored 3, because changes are possible, but since many equations and processes are involved, it is more complex to perform and track those changes.

PEF and SALCA obtained the best score in ‘Model and model results’ since they are easy to understand. Daisy and Animo scored 3 because the results are easy to interpret, but understanding the models requires more effort.

No ‘Authoritative body’ supports the models; thus, SALCA, Animo and Daisy scored 1 in the subcriterion. PEF scored 5 because the emission model used was recommended by European Commission, a well-trusted international body. SALCA, Daisy and Animo scored 3 in ‘Academic authority’ as national research institutions provide them, and PEF scored 1. All models also scored 5 in ‘Neutrality across industries, products or processes’ because they use an unbiased, objective methodology.

The models scored very similarly, with a difference of 8% in the total score. The comparison intended to show that many models can fit the LCA scope, but considering different purposes. Further work is needed through guidelines or other documents, in what situations they should be applied, and to force LCA practitioners to respect this adequacy as the scope of this study is to judge whether the models are suitable for LCA purposes in general.

Furthermore, when estimating and applying the emissions provided by the models in a case study, it is possible to identify the main differences and their effect on the impact categories in LCA when considering an entire system (e.g. machinery, water and fuel used).

3.2 Quantitative comparison: a case study of maize crop in Spain (temperate/mesothermal climate)

PEF, SALCA, Daisy and Animo were used to estimate nitrogen emissions due to the use of mineral fertilizers in an irrigated maize crop system in Spain. (Table 7). Approaches for Animo and Daisy’s calibration included adjusting, for instance, rates of photosynthesis, N uptake by the crop and N concentration in different plant organs. After calibration, the simulated crop yields in Daisy and Animo were only – 3% and – 4% of the observed yields, respectively, showing that the two models are able to produce reliable results for the system (Fig. 3).

None of the models estimated all parameters. The most worrying estimates not considered are denitrification in SALCA (possible overestimation of N₂O emissions could increase the impact on climate change), and NOₓ in SALCA, Daisy and Animo (possibly increasing impacts on photochemical ozone formation, particulate matter and marine eutrophication). N from water irrigation in PEF, and dry and wet deposition in PEF and SALCA should be considered in the future as they can contribute to more N as input into the system. Seed’s nitrogen supply was only considered in Daisy, but being 1% of the total input, it is not a significant loss for the other models in the present study. Irrigation was considered differently in Animo, Daisy and SALCA and is responsible for the 18% variation in N irrigation.

Animo estimated the highest nitrate leaching (43.7 kg N/ha/year) and PEF the lowest (17 kg N/ha/year). SALCA and PEF do not consider the evapotranspiration in the soil, directly affecting the estimated emissions. In addition, irrigation modelled in Daisy may be decreasing the actual value of nitrate leaching, especially compared with Animo, since in Daisy less irrigation went to the crop system. The variation in results for nitrate leaching was 61%. The loss of nitrate due to surface runoff estimated resulted in zero in Daisy and Animo.

SALCA estimated the highest NH₃ volatilization (3.7 kg N/ha/year) and Daisy and PEF the lowest (3.4 kg N/ha/year), varying by 8%, being the lowest variation between the emissions. Although SALCA considers direct and indirect forms of ammonia volatilization, Animo considers the fertilizer application technique, and Daisy and Animo take into account the dry and wet deposition of NH₄⁺ available in the air. Still, no significant difference was observed in the results.

SALCA is the only model that estimates NOₓ emissions, which means more impacts will be attributed to the system. However, this represents an advantage for the model in terms of coverage of nitrogen emissions.
PEF and SALCA estimate N₂O using EF, while Daisy and Animo consider the emission part from nitrification and another part from denitrification processes. In Daisy, the N₂O is directly estimated, but in Animo a fraction of 0.005 (fraction for loam soils in temperate climate regions from de Vries et al. (2003)) was applied to assume the amount of N₂O in total nitrification. In Animo and Daisy, a fraction of 0.02 (also from de Vries et al. (2003)) was applied to distinguish between N₂ or N₂O in the denitrification. For Daisy and Animo, N₂O emissions from denitrification are 0.134 and 0.002 kg N/ha/year. In summary, N₂O emissions (kg N-N₂O/ha/year) in PEF release 2.4, SALCA 1.6, Daisy 3.5 and Animo 2.9.

The N uptake applied in PEF was the same as in SALCA. That said, SALCA applied the average yield and a crop uptake coefficient for N uptake (13 kg N/ton DM). The variation in this output was 28%, 265.6 kg N/ha/year in SALCA (highest) and 190.3 kg N/ha/year in Daisy (lowest).

The highly negative N balance in PEF (−126.7 kg N/ha/year) is due to the limitation of N inputs considered and the ‘crop uptake’ being much higher than those estimated in Animo or Daisy. The balance in SALCA was −111.9 kg N/ha/year and did not consider N in soil and N mineralized as inputs into the system, although they have been used for NO₃⁻ leaching estimates. Again, crop uptake is a major contributor to the nigh negative balance in SALCA. Animo had the highest NO₃⁻ leaching output, resulting in an N balance of -59.3 kg N/ha/year. This high NO₃⁻ leaching was the distinguishing parameter that caused high N balance variation (59%) compared with the other mechanistic model Daisy, since other estimated emissions were similar. Daisy achieved the best balance (-29.3 kg N/ha/year) compared with the other models, considering that, although negative, is the closest to zero. According to the balances, there was a decrease in the soil mineral nitrogen stock.

Table 7 Average (2013–2017) nitrogen components estimated with the models PEF, SALCA, Daisy, Animo

| Source | PEF | SALCA | Daisy | Animo |
|--------|-----|-------|-------|-------|
| Input (kg N/ha/year) | Fertilizer (mineral fraction) | 170 | 170 | 170 | 170 |
| | Deposition | - | - | 15.6 | 14.7 |
| | Irrigation | - | 8.3 | 6.8 | 6.2 |
| | Plant N fixation | - | 0 | 0 | 0 |
| | N in soil | - | - | - | - |
| | Seed | - | - | 2.0 | - |
| | Total input | 170 | 178.3 | 194.4 | 190.9 |
| Output (kg N/ha/year) | Leaching to groundwater (N – NO₃⁻) | 17.0 | 18.0 | 19.9 | 43.7 |
| | Loss to surface water | - | - | 0 | 0 |
| | NH₃ Volatilization (N-NH₃) | 3.4 | 3.7 | 3.4 | 3.6 |
| | NO₂ | - | 1.2 | - | - |
| | Nitrification (N-N₂O) | 2.4 | 1.6 | 3.4 | 2.9² |
| | Denitrification (N-N₂O and N₂) | 7.65¹ | - | 6.7 | 0.7 |
| | N uptake | 265.6 | 265.6 | 190.3 | 199.0 |
| | Total output | 296.8 | 290.1 | 223.7 | 249.9 |
| | Balance | -126.7 | -111.8 | -29.3 | -59.0 |

¹N₂ emissions
²A fraction was used to separate N₂O emissions

Fig. 3 Calibration of Daisy and Animo models using the yield for irrigated maize in Spain
3.3 Characterization of impacts in an LCA of maize crop in Spain

The impacts were characterized in the Simapro software (Pré Consultants 2017), using a scenario provided by Montemayor et al. (2019) and the models’ emission estimates (Table 8). The impact categories analysed with the ILCD 2011 midpoint method were ‘Climate change (CC)’, ‘Particulate Matter (PM)’, ‘Photochemical Ozone Formation (POF)’, ‘Acidification (AC)’, ‘Terrestrial Eutrophication (TE)’ and ‘Marine Eutrophication (ME)’. Impact assessment models recommended in ILCD 2011 (EC-JRC 2011) midpoint method are available in Supplementary Material 4.

The impacts were calculated for 1 t of harvested maize dry matter (DM) (Table 9). Importantly, the variation in the values was caused only by the fertilizer emissions, since the ones related to machinery, fuels and other emissions were maintained the same as in Montemayor et al. (2019).

Although the impact variation among models was less than the variation in estimated emissions, the contribution of N from fertilizer input to impacts is evident. The 54% variation in N₂O emissions caused a 35% change in ‘CC’. The 9% variation in NH₃ emissions, caused a 1% change in the impact on ‘PM’ (smallest change in the calculated impacts), and an 18% change in ‘TE’. For ‘POF’, only SALCA provided NO₂ emissions, and these emissions caused a 31% change in impact. In ‘AC’, the NH₃ and NO₂ emissions caused a 14% change in the impact. The highest variation occurred in the impact category ‘ME’, with 56% change caused by a 61% variation in the NO₃⁻ leaching.

SALCA had the largest impacts on ‘PM’, ‘POF’, ‘AC’ and ‘TE’ due to the additional emissions of NO₂, in addition to the emission of NH₃. PEF emissions had the lowest impact in all impacts categories selected, except on ‘CC’.

A normalization procedure was carried out using the UE27 2010 methodology (Benini et al. 2014; Crenna et al. 2019) to compare the total impact and impact categories in the proposed scenarios (Fig. 4). Animo emissions caused the highest impact in the system, with a normalized score of 4.09, followed by SALCA (3.42), Daisy (3.29) and PEF (3.13), varying 23% in the normalized impact caused, only changing nitrogen emissions from fertilizer application. The models presented the same decreasing order of contribution in the impact categories: ‘PM’, ‘ME’, ‘TE’, ‘AC’, ‘CC’ and ‘POF’. However, the contribution of each impact category to the system is different. For instance, in ‘PM’, Daisy emissions contributed 25%, but in Animo the contribution was 20% of the total impact; in ‘ME’, 38% of overall impact was attributed to the NO₃⁻ leaching in Animo, but 22% in PEF. The different emissions directly affect the LCA final results, and this is also relevant when compared with other LCAs for irrigated maize crops or when calculating the system’s uncertainties.

4 Discussion

4.1 Comparison of model evaluation results with previous studies

Other studies comparing the models selected in this study have been performed (Wu and McGechan 1998; Cannavo et al. 2008; Bockstaller et al. 2009; Nitschelm et al. 2018; Peter et al. 2016) for various reasons and using different approaches.

Wu and McGechan (1998) compared Animo and Daisy (older versions) with two other mechanistic models (SOILN and SUNDIAL). Their results showed that Animo and Daisy have similarities, especially related to the effects of temperature and water content in the soil, but in denitrification

### Table 8 Fertilizer emissions used in a Spanish maize crop life cycle inventory for each N emission model, PEF, SALCA, Daisy and Animo

| N emission     | PEF  | SALCA | Daisy | Animo | Variation |
|----------------|------|-------|-------|-------|-----------|
| N₂O (kg N₂O/ha/year) | 3.8  | 2.5   | 5.5   | 4.6   | 54%       |
| NH₃ (kg NH₃/ha/year)  | 4.1  | 4.5   | 4.1   | 4.4   | 9%        |
| NO₂ (kg NO₂/ha/year) | -    | 5.3   | -     | -     | -         |
| NO₃⁻ (kg NO₃⁻/ha/year) | 75.3 | 79.7  | 88.1  | 193.5  | 61%       |

### Table 9 Impact characterization relevant to fertilizer emissions estimation using PEF, SALCA, Daisy and Animo models

| Impact category | Unit     | PEF     | SALCA   | Daisy   | Animo   | Variation |
|-----------------|----------|---------|---------|---------|---------|-----------|
| CC              | kg CO₂ eq/ton | 2669   | 2073    | 3175    | 2907    | 35%       |
| PM              | kg PM2.5 eq/ton | 4.17   | 4.22    | 4.17    | 4.19    | 1%        |
| POF             | kg NMVOC eq/ton | 8.42   | 12.26   | 8.42    | 8.42    | 31%       |
| AC              | molc H⁺ eq/ton | 25.60  | 29.62   | 25.60   | 26.51   | 14%       |
| TE              | kg N eq/ton   | 97.36  | 118.99  | 97.36   | 101.41  | 18%       |
| ME              | kg N eq/ton   | 21.00  | 23.53   | 23.90   | 47.74   | 56%       |

CC climate change, PM particulate matter, POF photochemical ozone formation, AC acidification, TE terrestrial eutrophication, ME marine eutrophication.
significant differences due to the applied parameters are present. They also pointed out that ammonia volatilization is modelled to a limited extent on both models, depending on the EF entered by the user. Agreeing Wu and McGechan (1998), denitrification in this study had an 89% difference between the Daisy and Animo estimates, and ammonia volatilization had only 8% of the difference between the models, being quite simplified even in the mechanistic models.

Cannavo et al. (2008) compared 62 mechanistic and empirical models, including Animo and Daisy, to assess environmental impacts of cultivated soils due to nitrogen emissions. Unlike this study, Cannavo et al. (2008) did not explain the simulated N processes. However, they pointed out that no lower performance was observed between empirical and mechanistic models, as long as the empirical models are applied in the specific context for which they were developed, respecting their geographic coverage, also spatial and temporal resolution required for the study’s goal and scope. In summary, they said that mechanistic and empirical models would provide different results due to the models’ internal parameters that were the same observed in the current work. The exception was for ammonia volatilization, in which all models obtained almost the same results, but this was expected since the models estimate ammonia volatilization simply and similarly.

Bockstaller et al. (2009) compared SALCA to three other models to test their capability as a farm management tool. SALCA obtained the best score for ‘environmental scientific soundness’ including coverage of agricultural production branches and coverage of production factors. However, SALCA was unable to cover all relevant environmental issues (e.g. biodiversity), and it was not considered user-friendly to farmers. Unlike the findings of Bockstaller et al. (2009), in the present study, SALCA is considered a user-friendly model compared with the mechanistic models, Daisy and Animo, but being related to the use by LCA practitioners.

Peter et al. (2016) and Torrellas et al. (2018) and compared Tier 1, Tier 2 and Tier 3 approaches to the estimation of greenhouse gases (GHG) in wheat crops and peach orchards, and emissions from a cow manure biogas plant in Catalonia, respectively. Both works used IPCC (2006) as Tier 1 model, Tier 2 model in Peter et al. (2016) was Bouwman et al. (2002) and in Torrellas et al. (2018) was regionalized EF to Catalonia. Regarding Tier 3 models, Peter et al. (2016) decided not to select any model justifying that, at the moment, there was no model readily available and easily implementable by the user, and Torrellas et al. (2018) used EF estimated from field measurement. Peter et al. (2016) found relevant differences in the estimates, up to + 50% between Tier 1 and Tier 2 models, similar to the current work (34%). In Torrellas et al. (2018) the difference between the results from Tier 1 and Tier 2 models were 24%, similarly obtained in the current work, and of 25% in average comparing Tier 1 and Tier 3 models, also similar to the 30% found in the present work. Both studies strongly recommended the use of higher Tier models to estimate nutrient emissions, and Peter et al. (2016) highlighted the convenient relation between reducing complexity and improving precision when using medium-effort (Tier 2 and Tier 3) models that is also expected and preferable to be applied in LCA studies.

Nitschelm et al. (2018) compared NO$_3^-$ and NH$_3$ emissions provided by a Tier 3 model Syst’N in a cropping system with the emissions estimated using the risk tables provided by AGRIBALYSE (Koch and Salou 2015), frequently used in LCA and similar to SALCA. For nitrate leaching, AGRYBALYSE models estimated emissions up to 67% lower than those estimated using Syst’N, similar to the differences found in this work, 58%, comparing SALCA and Animo’s results. Regarding NH$_3$, the differences were from 28 to 63%, thus higher than in the current work. In addition, the authors recommended Tier 2 and Tier 3 models for farming systems at regional scales, and Tier 1 models for more general assessments such as national environmental labelling of food products.
4.2 Comparison of simulation results provided by the models to field observations

The validation of the models' simulation against field measurements is essential to confirm if the results are accurate. However, due to the lack of field measurement for that specific system (which is common in LCA), the results of the simulation were compared with other studies containing similar environmental conditions and field practices.

The estimated values for N uptake in irrigated maize under Mediterranean conditions ranged from 151 to 254 kg N/ha (Berenguer et al. 2009), 262 to 333 kg N/ha (Yagüe and Quílez 2010) and 155 to 300 kg N/ha (Biau et al. 2012). Results obtained for N uptake in Daisy (190.3 kg N/ha/year), Animo (199.1 kg N/ha) and SALCA (265.6 kg N/ha) are in agreement with the interval found in field studies. Therefore, all models adequately estimated N uptake, despite the 28% variation in the estimated emissions.

In nitrate leaching Coefficient (NLC) in Mediterranean climate conditions, the interval for the nitrate leaching (kg N-NO₃/kg fertilizer applied) in irrigated maize crops was 0.11 to 0.37 (Lasa et al. 2011). Thus, SALCA (0.11), Daisy (0.12) and Animo (0.26) reached results similar to this value. The 0.10 in PEF is slightly below the minimum limit.

Bussink (1994) and Recio et al. (2018) observed rates of approximately 1.5% of total N applied using CAN ammonia volatilization under Mediterranean conditions. PEF (2.0%), SALCA (2.2%), Daisy (2.0%) and Animo (2.1%) reached rates very similar to those authors.

According to Cayuela et al. (2017), the general average EF N₂O (kg N-N₂O/kg N applied) for Mediterranean agriculture should be 0.005, being half of the value proposed by IPCC (0.01) and a quarter of the recommended value in EF (0.022). Cayuela et al. (2017) also proposed an EF for irrigated crops, 0.0063. Therefore, N₂O emissions in this work should be between 0.83 and 1.1 kg N-N₂O/ha. None of the models achieved these results, PEF and SALCA due to the EF applied. Daisy and Animo due to the uncertainty in N₂O emissions from nitrification and denitrification.

Denitrification calculated under Mediterranean climate in Teira-Esmatges et al. (1998) showed that (N₂O + N₂) losses represented 1.7% to 13.6% of the total N fertilizer applied. Therefore, the expected emissions between 2.89 and 23.21 kg N/ha were achieved by PEF and Daisy. The expected emissions values are summarized in Table 10.

4.3 Nitrogen emission models used in agricultural LCA studies

The use of IPCC (2006) EF for N₂O emissions appears to be the standard practice in LCA studies. However, as explained in Cayuela et al. (2017), the proposed factors are not adjusted for some climates, and the use of default EFs can result in erroneous emissions, as it happened with PEF and SALCA results. Mechanistic models also have been used to estimate N₂O emission in LCA, for instance, GREET 16 in Wang et al. (2007), DNDC in Goglio et al. (2014), and DAYCENT in Kim and Dale (2005). Although Animo and Daisy did not fall within the range of observed emission results, a better calibration of the models and an adjustment of internal parameters can be done, meaning that for N₂O emissions, Tier 3 models, such Daisy and Animo, could provide more adjusted estimates.

EF's use for NH₃ volatilization is widespread but from different sources other than IPCC (2006) used in PEF. Thomassen et al. (2008) and Xue et al. (2016) applied EF from previous studies that are more adjusted to climate conditions than IPCC (2006) EF. Tier 2 models, such as SALCA, are an excellent alternative for reducing complexity and improving precision for NH₃ volatilization. Tier 2 models were used in Mancuso et al. (2019), Romero-Gámez et al. (2014) and Wu et al. (2018), but a validation such as that carried out in this study is necessary.

Tier 3 models are more common for the NO₃⁻ leaching estimations. The complexity of the estimate can vary

| N parameter | Observed | PEF | SALCA | Daisy | Animo |
|-------------|----------|-----|-------|-------|-------|
| N uptake (kg N/ha) | 151–333¹ | - | 265.60 | 190.30 | 199.10 |
| NO₃⁻ leaching (NLC) | 0.11–0.37² | 0.10 | 0.11 | 0.12 | 0.26 |
| Volatilization (%) | ~ 1.5%³ | 2.0% | 2.2% | 2.0% | 2.1% |
| N₂O (kg N₂O emitted/ha) | 0.53–0.68⁴ | 3.8 | 2.5 | 5.5 | 4.6 |
| Denitrification (kgN/ha) | 2.89–23.21⁵ | 7.65 | - | 6.70 | 0.70 |
| Within the range | Not within the range | Not applicable |

¹Berenguer et al. (2009); Yagüe and Quílez (2010); Biau et al. (2012)
²Lasa et al. (2011)
³Bussink (1994) and Recio et al. (2018)
⁴Cayuela et al. (2017)
⁵Teira-Esmatges et al. (1998)
substantially under different climate conditions (i.e. dry and wet climate) and management operations (i.e. irrigation, free drainage and drainage with pipes). For example, DAYCENT in Kim and Dale (2005), DNDC in Goglio et al. (2014), STICS in Plaza-Bonilla et al. (2018) and Daisy, as aforementioned. Tier 1 models represented by different rates or EF have also been applied, for example, 0.25 for summer maize (Wang et al. 2007) and 0.26 for rice in Xue et al. (2016). Tier 2 models were applied in NO\textsubscript{3} leaching estimate in Romero-Gámez et al. (2014). For NO\textsubscript{3} leaching, Tier 3 models should be taken as first option to estimate this emission, since Tier 1 (PEF) and Tier 2 (SALCA) models may not be considering most parameters needed for a better estimate.

Usually, when authors use mechanistic models, all nitrogen emissions are estimated using the same model (Goglio et al. 2014; Kim and Dale 2005; Li et al. 2016; Plaza-Bonilla et al. 2018). The scientific advantage of using mechanistic models is the calibration performed, making the results more credible and appropriate to the system. For the validation in the aforementioned studies, literature data was used in Goglio et al. (2014) and Ni et al. (2019), as it was provided in this study. In Wang et al. (2007) validation was assumed from Hu (2004), another strategy that could be adopted in LCA. However, no validation of results can no longer be an option in LCA.

According to Nemecek et al. (2016), the ideal model should be practical; calculates the results easily, be site- and time-dependent (but to apply under a wide range of situations); and includes a collection of parameters and input data required. However, while no model complies all those important characteristics for LCA, mechanistic models, well-validated and calibrated for different situations, could be used to provide regionalized EF, as in Brown et al. (2002) and Yoshida et al. (2016), to be applied in lower Tier models to adjust N emissions.

5 Conclusions

PEF, SALCA, Daisy and Animo have important characteristics that make them useful and suitable for LCA, whenever their domains as fertilizer application emissions models are respected. Daisy was the model that best fitted to the criteria selected, achieving 77% of the total score. The proposed methodology could be used in other studies to compare models’ suitability for estimating nitrogen or nutrients in LCA.

For the case study applied, the models estimated reliable results for almost all N emissions, except for N\textsubscript{2}O. However, the characterization impact carried out showed differences in the impact categories analysed. Other crops should have their emissions estimated under different models to corroborate with the results in this work.

More research must go into emission model comparisons, describing more complex agricultural systems (including double crops, organic fertilizer including manure by-products, cultivation on substrates), to identify the best ways to estimate nitrogen emissions in LCA. Guidelines or methodologies are needed to guide the LCA practitioner to better describe and justify their agricultural inventory emissions choice. A sensitivity analyses that assess different models, literature values for similar crops, and field data could be used as a strategy to validate the results estimated.

Finally, it is not always possible to use mechanist models like Daisy or Animo to estimate nitrogen emissions in LCA, mostly due to the amount of input data required. However, after calibrations and validations, these models could be used to adjust EF, according to different climate conditions, crops and fertilizers used in the simplest models, such as SALCA or PEF. Therefore, LCA can benefit from using agricultural models, helping to improve their evidence-based results and recommendations.

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