Exposure to fogger trucks and breast cancer incidence in the Long Island Breast Cancer Study Project: a case–control study

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Abstract

Background: Few studies have supported an association between breast cancer and DDT, usually assessed with biomarkers that cannot discern timing of exposure, or differentiate between the accumulation of chronic low-dose versus acute high-dose exposures in the past. Previous studies suggest that an association may be evident only among women exposed to DDT during biologically susceptible windows, or among those diagnosed with estrogen receptor/progesterone receptor-positive (ER+PR+) breast cancer subtypes. Self-reported acute exposure to a fogger truck, which sprayed DDT prior to 1972, was hypothesized to increase the risk of breast cancer, particularly among women exposed at a young age or diagnosed with ER+PR+ breast cancer.

Methods: We examined these possibilities in the Long Island Breast Cancer Study Project (LIBCSP) (1,508 cases, 1,556 controls), which included exposure assessment by structured questionnaire and serum samples collected between 1996–1998, using adjusted logistic and polytomous regression to estimate ORs and 95% CIs.

Results: Women with ER+PR+ breast cancer had a 44% increased odds of ever seeing a pre-1972 fogger truck compared to other subtypes (OR = 1.44; 95% CI 1.08-1.93). However, there was little variation in the observed increase in breast cancer risk when considering all women who reported seeing a pre-1972 fogger truck at their residence (OR = 1.16; 95% CI 0.98, 1.37), or during hypothesized susceptible windows. Self-reported acute exposure was not correlated with serum concentrations, a biomarker of long-term exposure.

Conclusions: These findings support the hypothesis that seeing a fogger truck, a proxy measure for acute DDT exposure, may be associated with ER+PR+ tumors, the most commonly diagnosed breast cancer subtype among American women.

Keywords: Organochlorines, Dichlorodiphenyltrichloroethane (p,p′-DDT), 1,2-Dichloro-2,2-bis (p-chlorophenyl) ethylene (DDE), Pesticides, Cancer, Breast

Background

Breast cancer is the leading cancer diagnosis among women in the United States (U.S.) [1]. There is substantial evidence that the etiology of breast cancer may differ by receptor-defined subtypes [2]. The most commonly diagnosed subtypes among American women are luminal A and luminal B tumors [3], which together can be defined as estrogen and progesterone receptor positive (ER+PR+) tumors [4,5], and are often associated with estrogen-related risk factors such as early age at menarche and nulliparity [6-8].

There has long been suspicion of a relationship between the persistent organochlorine pesticide p,p′-dichlorodiphenyltrichloroethane (p,p′-DDT) and breast cancer, although results have been inconsistent [9]. By 1945, p,p′-DDT and its most prominent metabolite, p,p′-dichlorodiphenyldichloroethylene (p,p′-DDE), were widespread in the U.S. [10]. Fogger trucks sprayed DDT, the commercial product, on Long Island (Figure 1), and elsewhere across the nation, for the control of gypsy moths and mosquitoes until use greatly declined after being...
banned in 1972 (see timeline, Figure 2) [10]. \( p,p' \)-DDT exposure has been inconsistently linked to breast cancer despite having both estrogenic and carcinogenic properties [11,12]. Acute, high dose exposure to commercial DDT (which contains the more strongly estrogenic compounds \( o,p' \)-DDT and \( p,p' \)-DDT) has been shown to increase mammary cell proliferation in animal models [13,14] whereas chronic low-dose exposure through the diet is more predominately composed of the metabolite \( p,p' \)-DDE, which has little to no estrogenic effect [14]. DDT is still utilized to combat malaria, and more evidence regarding its role in possible health outcomes could either expand its use for vector control or encourage the use of other pesticides [15].

Early studies suggested that \( p,p' \)-DDT concentrations were associated with an increased risk of breast cancer [14,16]. However, most studies have supported little if any association [17–23]. Few studies have measured concentrations during developmental periods [24], which could represent a biologically plausible window of exposure for breast cancer risk [25,26], or among breast cancer subtypes. Instead, previous investigations often relied on biomarker assessments of \( p,p' \)-DDT metabolites. Positive correlations are observed with age and are likely due to the relatively long half-lives (more than 4 years for \( p,p' \)-DDT [27] and 10 years for \( p,p' \)-DDE [28]). Whether measured in fat or blood specimens, these biomarkers are therefore more likely to reflect cumulative, lifetime exposure. Thus, exposures during specific time periods, or acute versus chronic exposures, are not readily discernible, especially when these long-term biomarkers are assessed in blood samples collected long after the critical exposure of interest. Between 2000 and 2001, Cohn and colleagues [24] measured \( p,p' \)-DDT concentrations in serum samples collected beginning in 1959, the year of peak DDT use, and ending in 1967, when DDT was still in use. Among women who were younger than 14 in 1945, those with blood concentrations in the highest tertile were five times more likely to develop breast cancer compared to the lowest tertile. This is consistent with the hypothesis that breast cancer risk can be affected by events early in life [29], and there is interest at a national level regarding exposures that occur during adolescence and subsequent breast cancer risk [25].

The aim of this study was to investigate the relationship between self-reported exposure to a fogger truck at a residence prior to the ban of DDT, as a proxy measure of an acute high dose exposure, and breast cancer incidence, particularly among women diagnosed with ER+PR+ tumors. In addition, the association between adolescent exposure to fogger trucks during two demonstrated developmentally sensitive time periods, defined as \( \leq 20 \) years (consistent with John and Kelsey, 1993 [30]) and less \( \leq 14 \) years (consistent with Cohn et al., 2007 [24]), and breast cancer risk was determined. Further, the relationship between fogger truck exposure and breast cancer risk was assessed by birth cohort.

**Methods**

Data for this case–control investigation were drawn from the Long Island Breast Cancer Study Project (LIBCSP), which is described in detail in Gammon et al. [31] Briefly, LIBCSP is a population-based study of
breast cancer focusing on female residents in Nassau and Suffolk counties in Long Island, New York. IRB approval was obtained from all participating institutions.

Study population
All participants were residents of Nassau or Suffolk counties at the time of diagnosis and spoke English. Cases were women newly diagnosed with a first primary in situ or invasive breast cancer between August 1st, 1996 and July 31st, 1997, confirmed by a physician or the medical record. Cases were ascertained by contacting pathology departments of all 28 hospitals on Long Island and three tertiary care hospitals in New York City.

Controls were randomly selected women who had no history of breast cancer and were frequency matched to cases based on the expected age distribution of case women by 5-year age groups. Potential controls were identified by random digit dialing for those less than 65 years of age and by Health Care Finance Administration rosters for those 65 years and older.

Written signed informed consent was obtained from all participants prior to interview. A total of 1,508 cases and 1,556 controls (82.1% and 62.7%, respectively, of all eligible subjects) completed the interview process. Participants were women aged from 20 to 98 years, and most likely to be postmenopausal (67%); in addition, the overwhelming majority of women self-identified as white (93%), rather than as black (5%), which is consistent with the racial distribution of these two counties at the time of data collection [31].

Cases (97.7%) signed a medical record release form for the abstraction of clinical characteristics of breast cancer diagnosis and final study eligibility determination. For 95.2% of cases, medical records were successfully located and abstracted.

Assessment of fogger truck exposure
Exposure to fogger trucks was assessed by questionnaire. The LIBCSP case–control interview occurred in the respondents’ homes and, for cases, within several months of the diagnosis of the first primary breast cancer. All participants were administered the main case–control questionnaire by uniformly trained interviewers. The main questionnaire included a comprehensive assessment of known and suspected risk factors for breast cancer, including a section on environmental exposures. This section included a question asking whether a fogger truck was seen at a participant’s residence and if yes the ages or the dates at which they began and ceased living at the residence. Study participants were able to list 2 residences where they witnessed a fogger truck. Any other exposures to fogger trucks at non-residences would not be included in this question. Participants were also asked whether they chased a fogger truck during childhood or adolescence; however, due to low levels of a positive response that question was not included in this analysis.

To maximize information regarding windows of exposure, we utilized a birth cohort approach (see Figure 2). Specifically, the relationship between fogger truck exposure and breast cancer risk was assessed across 4 birth cohorts (<1925, 1925–1945, 1946–1972, >1972). Exposure at ≤20 years of age was defined as the hypothesized susceptible period, a time that encompasses both puberty and time before first birth for 90% of the LIBCSP population. The first (<1925) and last (>1972) birth cohorts would not have been exposed to DDT when ≤20 years of age. The birth cohort born between 1925 and 1945 would have been ≤20 years of age at the time of the introduction of DDT in 1945 [10] and those born in 1946–1972 would have been ≤20 years of age when DDT exposure was ubiquitous.

Other covariate assessment
Other LIBCSP data used for the current study include information collected as part of the in-person interview, from serum, or case medical record abstraction, as described below.

Responses to other sections of the main questionnaire, including reproductive and menstrual histories, life course body size, family history of breast cancer and race, were used when considering potential confounders and/or effect modifiers. Distributions of known risk factors among the study population have been described in full in Gammon et al. [31]. Serum were used to assess serum \( p,p' \)-DDT and \( p,p' \)-DDE concentrations. Certified phlebotomists or nurses obtained blood samples from 73.1% and 73.3% of cases and controls, respectively. Blood samples were randomly selected from participants who donated blood between 1996 and 1998, with a blood volume of greater than 1.5 ml, from all African-Americans and from all cases with \textit{in situ} breast cancer [17]. The Brock et al. [32] method was used to determine serum concentrations of \( p,p' \)-DDE and \( p,p' \)-DDT. Limits of detection were about 0.2 ng/ml for both \( p,p' \)-DDT and \( p,p' \)-DDE. \( p,p' \)-DDE serum concentrations were attained for 643 cases and 427 controls [17]. Blood samples were randomly selected from participants who donated blood between 1996 and 1998, with a blood volume of greater than 1.5 ml, from all African-Americans and from all cases with \textit{in situ} breast cancer [17]. The Brock et al. [32] method was used to determine serum concentrations of \( p,p' \)-DDE and \( p,p' \)-DDT. Limits of detection were about 0.2 ng/ml for both \( p,p' \)-DDT and \( p,p' \)-DDE. \( p,p' \)-DDE serum concentrations were attained for 643 cases and 427 controls [17].

Medical records of the cases were abstracted for 1,402 women to obtain data on hormone receptor status of the first primary breast cancer [17]. These data were used to classify cases by breast cancer subtypes. Information on hormone receptor status was available from the medical record for 990 cases (65.6%). As previously reported [17], among the LIBCSP breast cancer cases, the first primary tumor was ER+PR+ for 583 cases.
(58.9%), ER-PR- for 212 (21.4%), ER-PR+ for 52 (5.3%), and ER+PR- for 143 (14.4%).

**Statistical methods**

Descriptive statistics were determined for covariates, stratified by birth cohort and case status. Unconditional logistic regression [33] was used to estimate odds ratios (ORs) and 95% confidence intervals (CIs) for associations between breast cancer and reporting a fogger truck at a residence. The exposure was further stratified by whether the woman reported seeing a fogger truck ≤1972; if a woman reported seeing a fogger truck in both time periods, she was categorized in the ≤1972 group. Less than 10% of cases (n = 122) and controls (n = 124) reported seeing a fogger truck after the ban on DDT. Further analyses were restricted to those women who reported seeing a fogger truck ≤1972 and those who never reported seeing a fogger truck (n = 1365 cases and 1406 controls); estimates for the total LIBCSP population are also shown. There were women missing information on which residence they saw a fogger truck (n = 26 controls and n = 21 cases) and 4 women without sufficient information on ages of exposure to be classified who are not included in the sub analyses. To explore for possible age periods of increased susceptibility, two windows were considered: ≤14 and ≤20 years of age. In doing so, the exposed group was restricted to those women who reported seeing a fogger truck at the residence they began living ≤14 and ≤20 years of age, respectively. The association between the number of residences a participant reported seeing a fogger truck (defined as 0, 1 and 2) and breast cancer risk was determined. Disjoint categories of age were also investigated (≤14, 15–20, >20 years of age).

All statistical models were adjusted for the frequency matching factor, 5-year age group. Other potential confounders were identified (through a thorough review of the relevant literature and the use and analysis of a directed acyclic graph (DAG) [34]), and include: change in weight from age 20 to diagnosis (lost, stable, gained), menopausal status (pre, post), breastfeeding history (ever, never), family history of breast cancer in mother, sister or daughter (yes, no) and race (white, black, other). Confounding was evaluated by these covariates using backward selection from the full model identified by our causal diagram with a 10% change in estimate criterion. None of the covariates changed the estimate more than 10%, and thus only age group was included in the final models.

Polytomous logistic regression [35] was used to estimate the ratio of the ORs and 95% CI for associations between ever seeing a fogger truck at a residence when comparing ORs for breast cancer subtypes, as defined by combined ER/PR status (ER+PR+ vs. all other subtypes, and ER-PR- vs. all other subtypes). The associations between seeing a fogger truck during possibly susceptible periods, ≤14 years of age and ≤20 years of age, and breast cancer subtypes was also investigated.

Effect measure modification on a multiplicative scale was assessed by comparing the multivariable models with and without cross-product terms of the exposure and weight change [35]. To determine modification by weight gain, participants were stratified by whether they lost weight, maintained their weight within 3% in lbs [36], or gained weight from age twenty to age at reference (which was the date of the first primary breast cancer diagnosis for cases and the date of study identification for controls). Other cutpoints were considered, including a 10 kg change in weight (data not shown); the cutpoints used here were believed to best represent the data. Weight gain as a dichotomous variable (with stable weight plus weight loss as the referent) was also considered as a possible effect modifier, but no significant interaction was noted (data not shown). Effect measure modification was also evaluated for breastfeeding status (ever, never) and menopausal status (pre, post).

The analysis was further stratified by birth cohort, based on important years in the history of DDT use [10]. The association between seeing a fogger truck at a residence and breast cancer within each birth cohort was evaluated (<1925, 1925–1945, 1946–1972, >1972).

To assess correlation with serum concentrations and self-report of a fogger truck at a residence, a kappa statistic [37] was calculated with lipid-adjusted \( p,p'\text{-DDT} \) and \( p,p'\text{-DDE} \) concentrations dichotomized at the median. Linear regression was also employed to test whether self-report of a fogger truck predicted log-transformed \( p,p'\text{-DDT} \) or \( p,p'\text{-DDE} \) concentrations or the \( p,p'\text{-DDE} / p,p'\text{-DDT} \) ratio in the serum with adjustment for lipid concentrations [38], 5-year age groups, triglycerides, cholesterol, change in weight from age 20 to diagnosis, current BMI, breastfeeding history and race [33]. An interaction between current BMI and change in weight was investigated (data not shown) but was not a strong predictor and did not remain in the models. Positive and zero values of individual organochlorine concentrations below detection limit was set to the lowest positive value for that compound observed in samples.

All analyses were completed using SAS 9.2 (Cary, NC).

**Results**

The majority of the LIBCSP study population was born in the two birth cohorts that were exposed to DDT at young ages, 1925–1945 and 1946–1972 (Table 1). As would be expected, lactation history and menopausal status vary by birth cohort.

In the LIBCSP population, 589 (39.1%) and 558 (35.9%) cases and controls, respectively, reported ever seeing a
fogger truck at a personal residence on Long Island. When restricted to those who lived at their residences prior to the DDT ban in 1972, 446 (32.7%) and 408 (29.0%) cases and controls, respectively, reported seeing a fogger truck. As shown in Table 2, ever seeing a fogger truck at a residence was associated with little to no increase in risk of breast cancer (OR = 1.14; 95% CI: 0.98, 1.32) in the entire LIBCSP population. When the sample was restricted to the subset of women who lived at their residence ≤ 1972, seeing a fogger truck at a residence was similarly associated with a little to no increase in risk of breast cancer (OR = 1.16; 95% CI: 0.98, 1.37). Reporting seeing a fogger truck ever at a residence > 1972 was not associated with breast cancer risk (OR = 1.07; 95% CI: 0.94, 1.23), with 11.7% (n = 122) and 11.1% (n = 124) of cases and controls, respectively, reporting seeing a fogger truck after the DDT ban. Unless otherwise noted, all of the results presented here will be for those who reported seeing a fogger truck ≤ 1972.

In subgroup analyses, 11.6% of cases and 11.8% of controls were found to be at or under the age of 20 when living at the residence where they saw the fogger truck. Slightly fewer, 9.9% of cases and 10.9% of controls were living at the residence ≤ 14 years of age. The associations between fogger truck and breast cancer were not stronger when limiting the exposed to participants who reported seeing a fogger truck at a residence that they lived ≤ 14 years of age (OR = 0.99; 95% CI: 0.76, 1.28) and ≤ 20 years of age (OR = 1.08; 95% CI: 0.84, 1.37) (Table 2).

There was little to no increase in risk for reporting seeing a fogger truck at 2 residences (OR = 1.15; 95% CI: 0.82, 1.61) (Table 3). When using disjoint categories for age at residence, there were positive but imprecise associations for women who were 15–20 at the residence they saw a fogger truck (OR = 1.86; 95% CI: 0.98, 3.52). Due to limited sample sizes in these categories, the remaining subgroup analysis focuses on women who were ≤ 14 and ≤ 20 years of age at the reported residence.

### Table 1 Descriptive characteristics by birth cohort, LIBCSP, 1996-1997

| Birth cohorts<sup>a</sup> | 1925-1945 | 1946-1972 |
|-------------------------|-----------|-----------|
| <1925       | 1925-1945 | 1946-1972 |
| N = 465     | N = 1,587 | N = 1,008 |
| Cases       | Cases     | Cases     |
| Controls    | Controls  | Controls  |
| N (%)       | N (%)     | N (%)     |

#### Age at reference

| <35        | 0 (0.0)   | 0 (0.0)   |
| 35-44      | 0 (0.0)   | 0 (0.0)   |
| 45-54      | 0 (0.0)   | 0 (0.0)   |
| 55-64      | 0 (0.0)   | 0 (0.0)   |
| 65-74      | 111 (41.9)| 70 (35.0) |
| 75-84      | 134 (50.6)| 112 (56.0)|
| 85+        | 20 (7.5)  | 18 (9.0)  |

#### Household Income

| <20,000    | 95 (36.0) | 63 (31.7) |
| 20,000-34,999 | 112 (42.4)| 82 (41.2)|
| 35,000-49,999 | 28 (10.6)| 23 (11.6)|
| 50,000-69,999 | 16 (6.1) | 14 (7.0) |
| 70,000-89,000 | 7 (2.7)  | 7 (3.5)  |
| >90,000    | 6 (2.3)   | 10 (5.0)  |

#### Body Mass Index at referent

| <18.5     | <5<sup>b</sup> (1.5) | 9 (1.1)   |
| 18.5-24.9 | 99 (38.2) | 337 (42.4) |
| 250.29-99 | 97 (37.5) | 248 (31.2) |
| >30       | 59 (22.8) | 201 (25.3) |

#### Lactation

| Never     | 173 (65.3) | 114 (57.0) |
| Ever      | 92 (34.7)  | 86 (43.0)  |

#### Menopausal Status

| Pre-       | 0 (0.0)   | 0 (0.0)   |
| Post-      | 265 (100.0)| 200 (100.0)|

<sup>a</sup> Participants born after 1972 (n<5, all controls) were not displayed for confidentiality purposes.

<sup>b</sup> Cell is defined as <5 for confidentiality purposes.
Table 2 Association between fogger truck and breast cancer, stratified by time period and age at residence

| Population              | Self-reported fogger truck at residence | Cases N (%) | Controls N (%) | Age-adjusted OR (95% CI)* |
|-------------------------|----------------------------------------|-------------|----------------|--------------------------|
| Total LIBCSP population | Never                                  | 919 (60.9)  | 998 (64.1)     | 1.00 (reference)         |
|                         | Ever                                    | 589 (39.1)  | 558 (35.9)     | 1.14 (0.98, 1.32)        |
|                         | ≤20 years old                          | 161 (10.7)  | 168 (10.8)     | 1.04 (0.92, 1.18)        |
|                         | ≤14 years old                          | 135 (9.0)   | 155 (10.0)     | 0.99 (0.88, 1.13)        |
| Fogger truck ≤1972      | Never                                  | 919 (67.3)  | 998 (71.0)     | 1.00 (reference)         |
|                         | Ever                                    | 446 (32.7)  | 408 (29.0)     | 1.16 (0.98, 1.37)        |
|                         | ≤20 years old                          | 159 (11.6)  | 166 (11.8)     | 1.08 (0.84, 1.37)        |
|                         | ≤14 years old                          | 135 (9.9)   | 153 (10.9)     | 0.99 (0.76, 1.28)        |
| Fogger truck >1972      | Never                                  | 919 (88.3)  | 998 (88.9)     | 1.00 (reference)         |
|                         | Ever                                    | 122 (11.7)  | 124 (11.1)     | 1.07 (0.94, 1.23)        |

*Earliest self-reported fogger truck at residence among Long Island, New York women, Long Island Breast Cancer Study Project, 1996-1997. “Total LIBCSP population” includes both fogger truck exposures ≤1972 and >1972, “Fogger truck ≤1972” and “Fogger truck >1972” definitions are mutually exclusive categories. “Ever” includes exposure >20, ≤20 and ≤14 years of age; “≤20 years of age” includes exposures ≤14 years of age.

*Participants exposed (n=5) during susceptible windows not displayed for confidentiality purposes.

As shown in Table 4, women with ER+PR+ tumors had a significantly increased odds of ever seeing a fogger truck when compared to all other subtypes (OR = 1.44; 95% CI: 1.08, 1.93) when estimated as the ratio of the odds ratio using polytomous regression. In contrast, women who reported ever seeing a fogger truck did not have increased odds of having ER-PR- breast cancer compared to all other subtypes (OR = 0.91; 95% CI: 0.64, 1.29). When using a case-case analysis, comparing ER+PR+ to all other subtypes the OR was also significant (OR = 1.45; 95% CI: 1.10, 1.92). A pronounced, although imprecise, relationship with ER+PR+ tumors was observed for those who saw a fogger truck at a residence they lived at ≤20 (OR = 1.50; 95% CI: 0.97, 2.32) and ≤14 (OR = 1.15; 95% CI: 0.99, 1.33) years of age. Breast cancer tumor subtypes were also stratified by birth cohort, although study power was limited (see Additional file 1).

The covariate distributions for women known ER/PR status were very similar to women without known ER/PR status (data not shown). However, the effect estimate for the association for ever seeing a fogger truck and breast cancer was slightly higher among case women with known ER/PR status (OR = 1.33; 95% CI: 1.11, 1.59) than among all women regardless of ER/PR status (OR = 1.16; 95% CI: 0.98, 1.37), although with overlapping confidence intervals.

The OR for the association between exposure to a fogger truck and breast cancer was explored for effect modification by adult weight change, breastfeeding history and menopausal status at diagnosis as shown in Table 5. Although weight change was not a statistically significant modifier of the relationship between fogger trucks and breast cancer (LRT $\chi^2 = 1.86$, df = 1, p = 0.2), there was a slightly elevated association between women who gained weight and reported ever seeing a fogger truck at a residence compared to women who had gained weight but did not report seeing a fogger truck (OR = 1.21; 95% CI: 1.01, 1.45). Breastfeeding history and menopausal status were not found to be significant modifiers of the relationship between fogger trucks and breast cancer. However, a positive association for reporting a fogger truck and breast cancer risk was observed for postmenopausal women (OR = 1.24; 95% CI: 1.02, 1.51).

Study participants were categorized by their birth cohorts, defined based on important periods of DDT use (Table 6). An increased risk of breast cancer (OR = 1.29; 95% CI: 1.10, 1.52) was observed for women who were born between 1925 and 1945 and who reported seeing a fogger truck compared to women in the same birth cohort who did not see a fogger truck. There was a suggestion of an increased in breast cancer risk for seeing a fogger truck for women born before 1925 (OR = 1.35; 95% CI: 0.95, 1.89). No increase in risk was observed for women born after 1945.

The kappa statistic between self-report of ever seeing a fogger truck and lipid-adjusted blood serum concentrations of $p,p'$-DDT ($k = -0.062$) and $p,p'$-DDE ($k = -0.060$) did not indicate any agreement. Additionally, an adjusted linear regression model to predict log transformed $p,p'$-DDT and $p,p'$-DDE concentrations from self-report of a fogger truck at a residence did not identify an association (Table 7) ($p,p'$-DDT, $\beta = -0.0240$, SE = 0.0387; $p,p'$-DDE, $\beta = -0.1002$, SE = 0.0569; $p,p'$-DDT/ $p,p'$-DDE, $\beta = 1.5783$, SE = 4.6883).

**Discussion**

This is the first study to consider whether there is an association between reports of seeing a fogger truck and the risk of breast cancer. Among the LIBCSP population as a whole, there is little evidence that ever seeing a fogger truck was associated with an increase in risk of breast cancer, however, effect estimates were stronger among certain biologically susceptible subgroups. The strongest, most robust findings were observed for women with hormone receptor positive breast cancer.

Individuals with ER+PR+ tumors were 44% more likely to have reported ever seeing a fogger truck than were participants with other breast cancer subtypes. When restricted to those who were living at the residence ≤14 and ≤20 years of age, women with ER+PR+ tumors were similarly more likely to report seeing a fogger truck than other subtypes, ER+PR+ tumors, which include Luminal A and Luminal B tumors [5], are driven by estrogen and progesterone levels and most commonly associated with...
reproductive risk factors [8]. This is consistent with xenoestrogenic activity of commercial DDT [39]. Laboratory studies have found that p,p′-DDT enhances ER+PR+ tumor growth [40,41]. Therefore, it would be expected to observe the strongest effect of DDT exposure among women with this hormonally active subtype. A stronger risk of ER+PR+ tumors was not found in a previous LIBCSP report that focused on p,p′-DDT and p,p′-DDE serum concentrations, but we did not consider self-reported exposure to fogger trucks in that analysis [17].

ER+PR+ are the most common subtype of breast cancer in American women with approximately 70% of tumors classified as ER+PR+ or “hormonally responsive” [42]. The high prevalence of ER+PR+ breast cancer underscores the need to better understand the etiology of these specific tumors in an effort to improve public health risk reduction strategies in the U.S.

When stratifying by birth cohort, we found that seeing a fogger truck was most detrimental among women who were 20 or younger at the time DDT was widely introduced in the U.S in 1945. It is plausible that any effect of seeing a fogger truck after those early years was masked by the ubiquitous exposure to DDT. This result suggests that seeing a fogger truck had the most detrimental effect before exposure became omnipresent. The increased risk observed for postmenopausal women is likely due to the fact that women diagnosed with premenopausal breast cancer in 1996–1997 were less likely to be born in the earlier birth cohorts where an effect of DDT on breast cancer risk was observed.

The proxy measure of a fogger truck used in this study suggests that p,p′-DDT, rather than the metabolite p,p′-DDE, may be the carcinogenic compound; a result that is consistent with laboratory evidence and the Cohn et al., study [12,24]. This finding may also help to explain discrepancies in the previous literature that often relied on DDE measurements or did not investigate the association by tumor subtype. Although the heterogeneity in the ORs for the associations between seeing a fogger truck and breast cancer by change in weight was not significant, the study power for evaluating the interaction was low because of the small number of LIBCSP participants who maintained or lost weight. The elevated risk of breast cancer among those who gained weight and reported seeing a fogger truck at their residence suggests that DDT may have a stronger or more detectable effect when stored in adipose tissue of women who gain and store weight over time.

There was not a large difference in the results when >1972 reports of exposure to fogger trucks were excluded. However, due to the possible use of other pesticides after 1972 by fogger trucks it was important to consider possible exposure misclassification and to better isolate the effects of DDT exposure.

p,p′-DDT and p,p′-DDE serum concentrations, a surrogate measure of long-term exposures, were not correlated with seeing a fogger truck at a residence, a surrogate measure of acute exposures. This result is not unexpected given the long period from exposure to blood collection in this study, and evidence that current

| Table 3 Fogger truck and risk of breast cancer by number of residences and age at residencea |
| --- |
| **Population** | **Residential information** | **Self-reported fogger truck at residence** | **Cases N (%)** | **Controls N (%)** | **Age-adjusted OR (95% CI)b** |
| Total LIBCSP population | # of reported residences | 0 | 919 (60.9) | 998 (64.1) | 1.00 (reference) |
|  | 1 | 459 (30.4) | 423 (27.2) | 1.16 (0.99, 1.37) |
|  | 2 | 130 (8.6) | 135 (8.7) | 1.07 (0.83, 1.40) |
| **Age at residence** | None | 919 (60.1) | 998 (63.5) | 1.00 (reference) |
|  | >20 years old | 445 (29.1) | 403 (25.7) | 0.94 (0.72, 1.21) |
|  | 15-20 years old | 29 (1.9) | 15 (1.0) | 2.03 (1.08, 3.81) |
|  | ≤14 years old | 135 (8.8) | 155 (9.9) | 0.94 (0.72, 1.21) |
| Fogger truck ≤1972 | # of reported residences | 0 | 919 (68.9) | 998 (69.4) | 1.00 (reference) |
|  | 1 | 335 (25.1) | 367 (25.6) | 1.16 (0.97, 1.38) |
|  | 2 | 79 (5.9) | 73 (5.1) | 1.15 (0.82, 1.61) |
| **Age at residence** | None | 919 (64.7) | 998 (68.2) | 1.00 (reference) |
|  | >20 years old | 339 (23.9) | 297 (20.3) | 1.19 (0.99, 1.43) |
|  | 15-20 years old | 27 (1.9) | 15 (1.0) | 1.86 (0.98, 3.52) |
|  | ≤14 years old | 135 (9.5) | 153 (10.4) | 0.92 (0.72, 1.21) |

a Self-reported fogger trucks at residence among Long Island, New York women, Long Island Breast Cancer Study Project, 1996-1997, stratified by time period. Population definitions and exposure information are mutually exclusive categories. 
bAdjusted for 5-year age groups.
organochlorine serum concentrations may be decoupled from exposure levels due to intakes from food sources [43] and differing metabolism and degree of organochlorine storage in individuals [44]. As well, this measure of acute exposure is unlikely to be indicative of total DDT exposure levels as women were likely exposed to commercial DDT by sources outside their residences. This lack of correlation between serum concentrations and the fogger truck measure at least partially explains the discrepancies between the association found here and the previous LIBCSP publication focused on serum concentrations, which did not include the fogger truck exposure [17].

There was no increase in risk observed for reporting a fogger truck at multiple residences. Although the report of seeing a fogger truck may represent a single exposure, it is also possible that spraying occurred repeatedly at the residence which would make it difficult to discern exact exposure levels based on the questionnaire.

There was little evidence of an increased breast cancer risk for seeing a fogger truck with either of the two biologically susceptible windows investigated in this analysis, ≤14 and ≤20 years of age. Although cell sizes were small, slightly stronger estimates were observed among participants who reported seeing a fogger truck at their residence when they were 15–20 years of age.

The LIBCSP has extensive residential history information that allowed for the investigation of exposure during susceptible windows. A strength of this investigation is that the study cohort was alive during DDT use in the U.S., and was exposed at a range of ages. However, despite the relatively large, population-based sample size available, a larger sample would have provided a better opportunity to thoroughly investigate the subgroup analysis of biological windows of susceptibility. Additionally, due to the age of the study population, there were few premenopausal women who were within the hypothesized

| Population                        | Self-reported fogger truck at residence | Subtype | Cases (N) | Controls (N) | Age-adjusted ratio of the ORs (95% CI) |
|-----------------------------------|----------------------------------------|---------|-----------|--------------|---------------------------------------|
| Total LIBCSP population           | Ever                                    | Other subtypes | 778       | 1,556        | 1.00 (reference)                      |
|                                   |                                        | ER-PR-   | 212       | 1,556        | 0.98 (0.71, 1.35)                     |
|                                   |                                        | Other subtypes | 407       | 1,556        | 1.00 (reference)                      |
|                                   |                                        | ER+PR+   | 583       | 1,556        | 1.41 (1.08, 1.84)                     |
| ≤20 years of age                  | Other subtypes | 539       | 1,556        | 1.00 (reference) |
|                                   | ER-PR-       | 152       | 1,556        | 0.91 (0.71, 1.16) |
|                                   | Other subtypes | 298       | 1,556        | 1.00 (reference) |
|                                   | ER+PR+       | 393       | 1,556        | 1.25 (1.01, 1.54) |
| ≤14 years of age                  | Other subtypes | 526       | 1,556        | 1.00 (reference) |
|                                   | ER-PR-       | 146       | 1,556        | 0.84 (0.64, 1.11) |
|                                   | Other subtypes | 289       | 1,556        | 1.00 (reference) |
|                                   | ER+PR+       | 383       | 1,556        | 1.32 (1.05, 1.67) |
| Fogger truck ≤1972                | Other subtypes | 708       | 1,530        | 1.00 (reference) |
|                                   | ER-PR-       | 190       | 1,530        | 0.91 (0.64, 1.29) |
|                                   | Other subtypes | 373       | 1,530        | 1.00 (reference) |
|                                   | ER+PR+       | 525       | 1,530        | 1.44 (1.08, 1.93) |
| ≤20 years of age                  | Other subtypes | 538       | 1,528        | 1.00 (reference) |
|                                   | ER-PR-       | 151       | 1,528        | 0.89 (0.53, 1.49) |
|                                   | Other subtypes | 297       | 1,528        | 1.00 (reference) |
|                                   | ER+PR+       | 392       | 1,528        | 1.50 (0.97, 2.32) |
| ≤14 years of age                  | Other subtypes | 526       | 1,528        | 1.00 (reference) |
|                                   | ER-PR-       | 146       | 1,528        | 1.04 (0.87, 1.24) |
|                                   | Other subtypes | 289       | 1,528        | 1.00 (reference) |
|                                   | ER+PR+       | 383       | 1,528        | 1.15 (0.99, 1.33) |

*Earliest self-reported fogger truck at residence stratified by time period and age at residence among Long Island, New York women, Long Island Breast Cancer Study Project, 1996-1997. Populations are mutually exclusive categories; "Ever" includes exposure >20, ≤20 and ≤14 years of age; "≤20 years of age" includes exposures ≤14 years of age.

*For ER+PR+ comparison, "other subtypes" includes ER-PR-, ER+PR-, ER-PR+. For ER-PR- comparison, "other subtypes" includes ER+PR+, ER+PR-, ER-PR+.

*adjusted for 5-year age groups.
susceptible windows during the time periods where there was an effect observed for fogger truck exposure. Due to the case–control design of the study, we were not able to completely address age-period-cohort effects that would result in a more definitive case for critical windows of exposure. We were also unable to investigate any impact time spent outdoors or use of local produce that could have contributed to the understanding of this relationship.

Another potential limitation is differential recall bias. Cases who were concerned about their previous exposures to potentially harmful chemicals may have been more likely to report seeing a fogger truck. It is unlikely that this bias would differ by ER/PR status, yet it was when we assessed for potential heterogeneity among these subgroups that one of our strongest findings was observed. Approximately equal proportions of cases and controls reported that they believed that at least one environmental factor could cause breast cancer (69 percent and 68 percent, respectively) [45]; suggesting recall bias was not differential with respect to case–control status. It is difficult

Table 5 Fogger trucks and risk of breast cancer by weight change, breastfeeding and menopausal status

| Population                  | Modifiers | No fogger truck | Fogger truck | Age-adjusted OR (95% CI) | p-value |
|-----------------------------|-----------|-----------------|--------------|--------------------------|---------|
|                             |           | Cases (N)       | Controls (N) | Cases (N)                | Controls (N) |                 |               |
| Total population            | Weight change | Lost 78 109     | 29 42 1.00   | reference 0.04 (0.53, 1.68) | 0.3      |
|                             |           | Stable 80 75    | 32 34 1.00   | reference 0.73 (0.39, 1.39) |         |
|                             |           | Gain 814 761    | 528 482 1.00 | reference 1.19 (1.01, 1.40) |         |
| Breastfeeding status        | Never 612 633 | 416 370 1.00   | reference 1.17 (0.97, 1.40) | 0.7      |
|                             | Ever 307 365 | 173 188 1.00   | reference 1.08 (0.83, 1.40) |         |
| Menopausal status           | Pre-308 328 | 164 175 1.00   | reference 0.96 (0.73, 1.26) | 0.2      |
|                             | Post-592 633 | 414 357 1.00   | reference 1.25 (1.05, 1.50) |         |
| Fogger truck ≤1972          | Weight change | Lost 78 109     | 23 32 1.00   | reference 1.03 (0.54, 1.96) | 0.2      |
|                             |           | Stable 80 75    | 20 24 1.00   | reference 0.68 (0.32, 1.43) |         |
|                             |           | Gain 761 841    | 403 352 1.00 | reference 1.21 (1.01, 1.45) |         |
| Breastfeeding status        | Never 612 633 | 311 273 1.00   | reference 1.16 (0.95, 1.43) | 0.9      |
|                             | Ever 307 365 | 135 135 1.00   | reference 1.13 (0.85, 1.51) |         |
| Menopausal status           | Pre-308 328 | 100 102 1.00   | reference 0.97 (0.70, 1.35) | 0.3      |
|                             | Post-592 633 | 339 292 1.00   | reference 1.24 (1.02, 1.51) |         |

aEarliest self-reported fogger truck at residence stratified by time period among Long Island, New York women, Long Island Breast Cancer Study Project, 1996-1997. Population and modifier definitions are mutually exclusive categories.
bAdjusted for 5-year age groups.
cMultiplicative interaction p-value from likelihood ratio test.

Table 6 Association between fogger truck and breast cancer risk, stratified by birth cohort

| Birth cohort | Population | Self-reported fogger truck at residence | Cases N (%) | Controls N (%) | Age-adjusted OR (95% CI) | p-value |
|--------------|------------|----------------------------------------|-------------|----------------|--------------------------|---------|
| <1926        | Total Population | No fogger 167 (63.0) | 138 (69.0) | 1.00 reference |               |
|              |            | Fogger 98 (37.0)         | 62 (31.0)  | 1.28 (0.86, 1.89) |     |
| Fogger truck ≤1972 | No fogger 167 (66.5) | 138 (72.6) | 1.00 reference |               |
|              |            | Fogger 84 (33.5)         | 52 (27.4)  | 1.35 (0.95, 1.89) |     |
| 1926-1945    | Total Population | No fogger 434 (55.6) | 509 (63.1) | 1.00 reference |               |
|              |            | Fogger 346 (44.4)        | 298 (36.9) | 1.36 (1.11, 1.66) |     |
| Fogger truck ≤1972 | No fogger 434 (61.0) | 509 (67.7) | 1.00 reference |               |
|              |            | Fogger 277 (39.0)        | 243 (32.2) | 1.29 (1.10, 1.52) |     |
| 1946-1972    | Total Population | No fogger 318 (68.7) | 348 (63.9) | 1.00 reference |               |
|              |            | Fogger 145 (31.3)        | 197 (36.2) | 0.80 (0.61, 1.04) |     |
| Fogger truck ≤1972 | No fogger 318 (78.9) | 348 (75.5) | 1.00 reference |               |
|              |            | Fogger 85 (21.1)         | 113 (24.5) | 0.87 (0.72, 1.04) |     |

aEarliest self-reported fogger truck at residence stratified by time period among Long Island, New York women, Long Island Breast Cancer Study Project, 1996-1997. Birth cohorts and population definitions are mutually exclusive categories.
bAdjusted for 5-year age groups.
to hypothesize at what age seeing a fogger truck could be considered a memorable experience and likely to vary across individuals. Any variation in recall would not be expected to differ by case status and thus would bias estimates towards the null. To the best of our knowledge, the use of any other pesticides in fogger trucks before the DDT ban is unknown and any exposure misclassification resulting from the use of alternative pesticides would attenuate any effect on breast cancer risk, which we assume is attributable to DDT. Alternatively, it is also possible that our results could be due to an unknown chemical or chemicals and not attributable to DDT.

We did not observe significant heterogeneity when investigating breast cancer risk by birth cohort (data not shown) and therefore it is unlikely that truck sightings are proxies for age at diagnosis. From the lack of an increase in risk observed for seeing a truck among the 1946–1972 birth cohort, we can also conclude that seeing a fogger truck was not a proxy for being at a young and potentially susceptible age during DDT use.

Few studies have examined the association between DDT exposure during younger ages and subsequent breast cancer risk, and none have considered acute DDT exposures. Therefore, the study results reported here provide a novel perspective to the ongoing discussion about the relationship between DDT and breast cancer. The stronger association observed with a biologically susceptible subgroup, namely women with ER+PR+ breast cancer, provide support for an association between DDT and breast cancer risk. With the prevalence of DDT increasing with indoor residual spraying programs for malaria vector control outside the U.S. [46], the importance of understanding the health effects of DDT continues.

Conclusions

In this first report that considers acute exposure to DDT, a positive association was observed for spraying fogger trucks and ER+PR+ breast cancer. Given the high incidence of breast cancer, and in particular ER+PR+ tumors [42], this finding suggests that increased risks for tumor subgroups may have significant public health importance in areas where DDT continues to be used. Although the potential association between DDT and breast cancer has been extensively addressed in previous studies, results are inconsistent [9]. The novel approach undertaken here, which focuses on the associations with seeing a fogger truck as a proxy for acute DDT exposure, provides new information regarding the relationship with ER+PR+ breast cancer subtype. Therefore, these findings, based on consideration of exposure timing and breast cancer subtype, may help to guide future investigations on the potential impact and carcinogenic mechanisms of DDT and breast cancer.

Additional file

Additional file 1: Breast cancer subtype, stratified by birth cohorts, LIBCSP, 1996-1997.

Abbreviations

ER+PR+: Estrogen receptor/ progesterone receptor-positive; p,p’-DDT; p,p’-dichlorodiphenyltrichloroethane; p,p’-DDE: p,p’-dichlorodiphenyldichloroethylene; EPA: U.S. Environmental Protection Agency; LIBCSP: Long Island Breast Cancer Study Project.

Competing interests

The authors declared that they have no competing interests.

Authors’ contributions

AW, ST and MG conceived and designed the analysis. AW and MG drafted the manuscript. AW carried out the statistical analysis. MW completed the organochlorine assays. MG was Principal Investigator, AN was co-Principal Investigator, and ST, MW and SS were collaborators on the LIBCSP, and thus were responsible for all data collection (study design and planning, including subject identification, recruitment, interview administration and phlebotomy). All authors contributed to the revisions of the manuscript and read and approved the final manuscript.

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References

1. American Cancer Society: Cancer Facts and Figures 2011. Atlanta, GA: American Cancer Society; 2011.
Xenoestrogens alter mammary gland morphology and breast cancer risk. *Oncologist* 2010, 15(Suppl 7):157–65.

6. Phipps AI, Malone KE, Porter PL, Daling JR, Li CI: Reproductive and hormonal risk factors for postmenopausal luminal, HER-2-overexpressing, and triple-negative breast cancer. Cancer 2008, 113:1521–1526.

7. Yang XR, Chang-Claude J, Goode EL, Yang XR, Chang-Claude J, Goode EL, Couch FJ, Nevalinna H, Milne RL, Gaudet M, Schmidt MK, Broaks A, Cox A, Fasching PA, Hein R, Spurdle AB, Blows F, Driver K, Velez-Jaque D, Heinz J, Sinn P, Viteli A, Heikkinen T, Attomomki H, Heikkila P, Biemniod C, Lisowski J, Peplonska B, Chanoq S, Figueroa J, et al: Associations of breast cancer risk factors with tumor subtypes: a pooled analysis from the Breast Cancer Association Consortium studies. *J Natl Cancer Inst* 2011, 103:250–263.

8. Yang XR, Sherman ME, Rimm DL, Lisowski J, Britton LA, Peplonska B, Hewitt SM, Anderson WF, Szenza-Dabrowska N, Bardin-Mikolajczak A, Zatonski W, Cartun R, Mandich D, Rymkiewicz G, Ligaj M, Lukaszek S, Kordok R, Garcia-Closas M: Differences in risk factors for breast cancer molecular subtypes in a population-based study. *Cancer Epidemiol Biomarkers Prev* 2007, 16:439–443. doi:10.1158/1055-9965.EPI-06-0806.

9. Lopez-Cervantes M, Torres-Sanchez L, Tobias A, Lopez-Carrillo L: Dichlorophenyl dichloroethane burden and breast cancer risk: a meta-analysis of the epidemiologic evidence. *Environ Health Perspect* 2004, 112:207–214.

10. US Environmental Protection Agency (EPA): DDT, a Review of Scientific and Economic Aspects of the Decision to Ban Its Use as a Pesticide. Washington, DC: US Environmental Protection Agency; 1975.

11. Wolff MS, Toniolo PG: Environmental organochlorine exposure as a potential etiologic factor in breast cancer. *Environ Health Perspect* 1995, 103(Suppl 1):141–145.

12. International Agency for Research on Cancer (IARC): DDT and associated compounds. Lyon, France: IARC: Monographs on Evaluation of Carcinogenic Risk to Humans; 1991.

13. Brown NW, Lamartiniere CA: Xenoestrogens alter mammal gland differentiation and cell proliferation in the rat. *Environ Health Perspect* 1995, 103:209–217.

14. Snedeker SM: Pesticides and breast cancer risk: a review of DDT, DDE, and dieldrin. *Environ Health Perspect* 2001, 109(Suppl 1):35–47.

15. Bouwman H, van den Berg H, Kyllin H: DDT and malaria prevention: addressing the paradox. *Environ Health Perspect* 2011, 119:744–747.

16. Wolff MS, Toniolo PG, Lee EW, Rivera M, Dubin N: Blood levels of organochlorine residues and risk of breast cancer. *J Natl Cancer Inst* 1993, 85:648–652.

17. Gammon MD, Wolff MS, Neugut AI, Eng SM, Teitelbaum SL, Britton JA, Terry MB, Levin SM, Belland SD, Kabat GC, Hatch M, Senie R, Berkowitz GS, Bradlow HL, Garbowski G, Maffeio C, Montalvan P, Kemeny M, Citron M, Schnabel F, Schuss A, Hadiju S, Vincquevara V, Niquidula N, Ireland K, Santella RM: Environmental toxins and breast cancer on Long Island. II. Organochlorine compound levels in blood. *Cancer Epidemiol Biomarkers Prev* 2011, 20:1168–1176.

18. Brody JG, Aschenbraun A, McKelway W, Rudel RA, Swartz CH, Kennedy T: Breast cancer risk and historical exposure to pesticides from wide-area applications assessed with GIS. *Environ Health Perspect* 2004, 112:889–897.

19. Charlier C, Foldi JM, Pitsane F, Hexman P, Gasgaard U, Meurisse M, Piomontu G: Environmental dichlorodiphenyltrichloroethane or hexachlorobenzene exposure and breast cancer: is there a risk? *Clin Chim Acta* 2004, 42:222–227.

20. Rubin CH, Lanier A, Kiesak S, Brock JW, Koller KK, Strosnider H, Needham L, Zahn S, Harpster A: Breast cancer among Alaska Native women potentially exposed to environmental organochlorine chemicals. *Int J Circumpolar Health* 2006, 65:18–27.

21. Gatto NM, Longnecker MP, Press MF, Sullivan-Halley J, Mc Keen-Cowdin R, Bernstein L: Serum organochlorines and breast cancer: a case–control study among African-American women. *Cancer Causes Control* 2007, 18:29–39.

22. Iwasaki M, Inoue M, Sasazuki S, Kurihashi N, Itoh H, Usuda M, Tsugane S, Japanese Public Health Center-based Prospective Study Group: Plasma organochlorine levels and subsequent risk of breast cancer among Japanese women: a nested case–control study. *Sci Total Environ* 2008, 402:176–185.

23. Siddiqui MK, Anand M, Mehrotra PK, Saniangi R, Mathur N: Biomonitoring of organochlorines in women with benign and malignant breast disease. *Environ Res* 2005, 98:250–257.

24. Cohn BA, Wolff MS, Critillo PM, Sholdt RL: DDT and breast cancer in young women: new data on the significance of age at exposure. *Environ Health Perspect* 2007, 115:1406–1414.

25. Hatt RA, Haslam SZ, Ousch J, Breast Cancer and the Environment Research Centers: The breast cancer and the environment research centers: transdisciplinary research on the role of the environment in breast cancer etiology. *Environ Health Perspect* 2009, 117:1814–1822.

26. Cohn BA: Developmental and environmental origins of breast cancer: DDT as a case study. *Reprod Toxicol* 2011, 31:302–311.

27. Longnecker MP: Invited Commentary: Why DDT matters now. *Am J Epidemiol* 2005, 162:726–728.

28. Wolff MS, Berkowitz GS, Brower S, Senie R, Bleiweiss IJ, Tartter P, Prince B, Roy N, Wallenstein S, Weston A: Organochlorine exposures and breast cancer risk in New York City women. *Environ Res* 2000, 84:151–161.

29. Goldstein GA, Frazier AL, Mink-Jones MJ: Levels of breast cancer show that risk is set by events of early life: prevention efforts must shift focus. *Cancer Epidemiol Biomarkers Prev* 2005, 14:567–571.

30. John EM, Kelsey JL: Radiation and other environmental exposures and breast cancer. *Epidemiol Rev* 1993, 15:157–162.

31. Gammon MD, Neugut AI, Santella RM, Teitelbaum SL, Britton JA, Terry MB, Eng SM, Wolff MS, Stelland SD, Kabat GC, Levin B, Bradlow HL, Hatch M, Beyea J, Camann D, Trent M, Senie RT, Garbowski GC, Maffeio C, Montalvan P, Berkowitz GS, Kemeny M, Citron M, Schnabel F, Schuss A, Hadiju S, Vincquera V, Collman GW, Obrams GI: The Long Island Breast Cancer Study Project: description of a multi-institutional collaboration to identify environmental risk factors for breast cancer. *Breast Cancer Res Treat* 2002, 74:235–254.

32. Brock W, Burse WV, Ashley DL, Naja’a AR, Green VE, Koner MP, Powell MK, Hodge CC, Needham LL: An improved analysis for chlorinated pesticides and polychlorinated biphenyls (PCBs) in human and bovine sera using solid-phase extraction. *J Anal Toxicol* 1996, 20:504–508.

33. Selvin S: Statistical Analysis of Epidemiologic Data. Second edition. New York: Oxford University Press; 1996.

34. Shrier I, Platt RW: Reducing bias through directed acyclic graphs. *BMC Med Res Methodol* 2003, 8:170.

35. Hosmer D, Lemeshow S: Applied Logistic Regression. New York: John Wiley & Sons; 1989.

36. Stevens J, Truesdale KP, McClain JE, Cai J: The definition of weight maintenance. *Int J Obes (Lond)* 2006, 30:391–399.

37. Cohen J: A coefficient of agreement for nominal scales. *Educ Psychol Meas* 1960, 20:33–45.

38. Phillips DL, Prikle JL, Burse WV, Bernett JT, Jr, Henderson LO, Needham LL: Chlorinated hydrocarbon levels in human serum: effects of fasting and feeding. *Arch Environ Contam Toxicol* 1989, 18:495–500.

39. Sheikhar PV, Werdell J, Basur V: Environmental estrogen stimulation of growth and estrogen receptor function in preneoplastic and cancerous human breast cell lines. *J Natl Cancer Inst* 1997, 89:1774–1782.

40. Robinson AK, Srirabuk DA, Stancel GM: DDT supports the growth of an estrogen-responsive tumor. *Toxicol Lett* 1985, 27:109–113.

41. Scribner JD, Mottet NK: DDT acceleration of mammary gland tumors induced in the male Sprague–Dawley rat by 2-acetamidophenanthrene. *Carcinogenesis* 1981, 2:1235–1239.

42. Potter JD, Cerhan JR, Sellers TA, McGovern PG, Orinkad C, Kushki LR, Lores AR: Progestosterone and estrogen receptors and mammary neoplasia in the Iowa Women’s Health Study: how many kinds of breast cancer are there? *Cancer Epidemiol Biomarkers Perspect* 1995, 4:319–326.

43. Longnecker MP, Rong WN, Lucier G: The human health effects of DDT (dichlorodiphenyltrichloroethane) and PCBs (polychlorinated biphenyls) and an overview of organochlorines in public health. *Annu Rev Public Health* 1997, 18:211–244.
44. Sholtz RI, McLaughlin KR, Cirillo PM, Petreas M, Park JS, Wolff MS, Factor-Litvak P, Eskenazi B, Krigbaum N, Cohn BA: Assaying organochlorines in archived serum for a large, long-term cohort: implications of combining assay results from multiple laboratories over time. Environ Int 2011, 37:709–714.

45. Teitelbaum SL, Gammon MD, Britton JA, Neugut AI, Levin B, Stellman SD: Reported residential pesticide use and breast cancer risk on Long Island. New York. Am J Epidemiol 2007, 165:643–651.

46. Van den Berg H: Global status of DDT and its alternatives for use in vector control to prevent disease. Environ Health Perspect 2009, 117:1656–1663.

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