Response of soil microbiota to nine-year application of swine manure and urea

Resposta da microbiota do solo a nove anos de aplicações de dejetos suínos e ureia

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ABSTRACT

Manure fertilization is a common practice, but little is known about its impacts on soil microbial activity and organic matter. Aiming to evaluate soil microbial response to nine years of successive applications of swine manure, organic carbon (TOC), total nitrogen (TN), pH, microbial biomass carbon (MBC), basal respiration (BR), metabolic quotient (qCO₂), and enzyme (β-glucosidase, phosphatase, arylsulphatase, and FDA) activities were measured in the 0-10cm soil layer, in a no-tillage system. Treatments were: control soil without fertilization (C), and application of two doses (104 and 209kg of N ha⁻¹year⁻¹) of urea (U₁ and U₂), pig slurry (PS₁ and PS₂) and deep litter (DL₁ and DL₂). TOC, TN, soil pH, MBC, and BR increased in soil fertilized with DL, and were lower in U treatments. Soils with U and DL application had higher qCO₂, related to different sources of stressors like nutrient imbalance. Phosphatase and β-glucosidase activities were not affected by treatments, increased with time, and had a strong correlation with MBC. We conclude that long-term swine manure applications increase microbial activity and soil organic matter, mainly in DL form; while urea applications have negative impacts on these indicators.

Key words: soil microbial activity, nitrogen amendments, soil quality.

INTRODUCTION

Confined pork production generates large amounts of manure in small areas, and waste matter is stored and commonly used for crop fertilization. Use of swine manure as a fertilizer promotes nutrient recycling and improves soil structure and fertility, but mismanagement may result in environmental contamination because agricultural areas for disposal are limited. Thus, repeated applications may result in nutrient leaching, producing water contamination, and promoting algal blooms and eutrophication in surface water (ZHANG et al., 2012).
Attributes of soil microbial community reflect effects of agricultural practices, like fertilization, because they respond quickly to changes in the environment, regulating organic matter decomposition and nutrient transformations. The most common biological soil quality indicators are: soil microbial biomass, basal respiration and metabolic quotient, the ratio between the former two parameters. Enzymatic activity is also commonly used to estimate activities related to specific nutrient cycles (GIACOMETTI et al., 2013).

Soil microbial biomass, consisting mainly of fungi and bacteria, is responsible for almost all-biological transformation in the soil. Since changes in microbial biomass are observed on a daily scale, while changes in soil organic matter may take decades, microbial biomass is used to assess the impact of different sources of stress on microorganisms. However, it must be evaluated along with others variables to assess soil biological processes (ANDERSON & DOMSCH, 2010).

Basal respiration is an indicator of biological activity, especially catabolic processes of aerobic microorganisms related to organic matter decomposition, and to nutrient release or immobilization in the soil. Higher basal respiration can indicates enhanced microbial activity, but it can also indicate loss of soil organic matter especially in situations where C input is smaller than C output (ANDERSON & DOMSCH, 2010). The ratio of basal respiration to microbial biomass, metabolic quotient ($q_{CO_2}$), is used to evaluate microbial efficiency for substrate conversion in cell mass. High $q_{CO_2}$ values mean low growth efficiency, since C lost as CO$_2$ is not used in new cellular tissue formation, also can mean stress due to biotic or abiotic factors and management conditions (WARDLE & GHANI, 1995). Soil enzymes, synthesized mainly by microorganisms, have an important role in nutrient cycle, and their activities are responsive to changes in soil and plant management, such as residue or fertilizer application, as well as to crop rotation (GIACOMETTI et al., 2013).

Comparisons between organic and mineral fertilizers have shown that microbial communities respond differently to distinct nutrient sources. Soil biomass, CO$_2$ evolution and enzymatic activity have shown increases above 30% in areas fertilized with swine manure in temperate regions (e.g., PLAZA et al., 2007; GIACOMETTI et al., 2013). On the other hand, mineral fertilizer application results in lower microbial biomass (above 20%), C/N ratio and enzyme activity, mainly β-glucosidase and phosphatase (MARSCHNER et al., 2003; GIACOMETTI et al., 2013). Furthermore, there is abundant literature describing pig slurry effects, but fewer studies have focused on swine manure in solid form (as deep litter) and compared to urea applications.

Thus, the present research aimed to evaluate changes in the response of soil microorganisms after nine years of repeated application of different kind of fertilizer, in a sandy Typic Haplustalf soil under no-tillage management. Particularly, we aimed to compare swine manure fertilization, in liquid (PS) and solid (DL) forms, as well as urea, evaluating how those fertilizers affected biological indicators of soil quality.

**MATERIAL AND METHODS**

The research was performed in a pork production farm in Southern Brazil (28°14’20.7"S and 49°13’55.5"W, 300m elevation) with Cfa humid subtropical climate and a sandy Typic Haplustalf soil (SOIL SURVEY STAFF, 2014). Before the establishment of the experiment (2002), physical and chemical attributes in 0-10 cm layer were measured: sand 330g kg$^{-1}$; organic matter 33g kg$^{-1}$; pH (water) 5.1; pH (SMP solution) 5.5; Mehtlich-1 extractable P 19.3mg dm$^{-3}$; KCl 0.1mol L$^{-1}$ exchangeable Mg 0.8cmol c kg$^{-1}$; KCl 0.1mol L$^{-1}$ exchangeable Ca 3.1cmol kg$^{-1}$; KCl 0.1mol L$^{-1}$ exchangeable Al 0.8cmol kg$^{-1}$; KCl 0.1mol L$^{-1}$ exchangeable Mg 0.8cmol kg$^{-1}$.

The experimental plots were established in 2002 with the following treatments: control without fertilization (C); two doses of annual pig slurry fertilization (PS, and PS$^1$), deep-litter fertilization (DL, and DL$^1$) and urea fertilization (U, and U$^1$). The doses were an average of 104 and 209kg N ha$^{-1}$year$^{-1}$ respectively, taking in account the total N content of each fertilizer. The experimental design was in completely randomized blocks with three replications, and each plot was 27m$^2$ (4.5×6.0m). PS was a mixture of swine excreta, cleaning water and feed remains; DL contained wood shavings, scraps and partially decomposed pig manure. The organic fertilizers and U were applied annually on the soil surface, without incorporation into the ground (Figure 1).

Soil samples were collected every 60 days from October 2010 to October 2011. Each soil sample consisted of 15 subsamples taken randomly from the 0-10cm layer with an auger. After, soil samples were packed and transported in containers with ice. Part of the soil was air dried and set aside for chemical analyses; the other part was stored at 4°C until the microbiological assays were performed. Total nitrogen (TN) was evaluated by the Kjeldahl method, total organic carbon (TOC) by the Walkley-Black method, and pH in water, as described in TEDESCO et al., 2013.}

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et al. (1995). Microbial biomass carbon (MBC) was determined by fumigation - extraction method, and prior to measurements the soil was incubated at room temperature and at moisture equivalent to 60% of total pore volume. Basal respiration (BR) was measured using the closed jar method. The air-tight jars contained 20 g of soil and were incubated for six days, at room temperature and darkness. Then, one beaker with 20mL of NaOH 0.2mol L\(^{-1}\) was placed in each jar, and subsequently, remaining NaOH was determined by titration with HCl 0.2mol L\(^{-1}\). Metabolic quotient (\(q_{\text{CO}}\)) was calculated as the ratio between basal respiration and microbial biomass (C-\(\text{CO}_2\)/MBC). Activities of the enzymes \(\beta\)-glucosidase, acid phosphatase, arylsulphatase and hydrolysis of fluorescein diacetate (FDA) were measured using colorimetric methods. All of these methods were made following the described by ALEF & NANNIPIERI (1995). Sampling dates are shown in figure 1.

RESULTS AND DISCUSSION

Soil pH was lower in the U and PS\(_2\) areas, as compared to other treatments (Table 1). Soil pH decreased with successive application of fertilizers, a process enhanced in soils receiving urea or materials containing urine, such as PS, due to the protons release during the oxidation of N; on the other hand, higher carbonate content in DL promotes consumption of H\(^+\) dissolved in the soil, leading to pH increases (BRUNETTO et al., 2012). In this study, pH of DL treatment was similar to the control soil, with values around 6.0. Lower pH in the U and PS\(_2\) treatments may decrease P uptake by plants (MEURER, 2007).

Areas receiving DL had increases of soil TOC and TN (up to 58 e 70%, respectively), in relation to control treatment (Table 1). PS and U showed lower contributions on the soil TOC and TN content. DL has higher contents of organic matter and recalcitrant carbon compounds with low mineralization rates (GIACOMINI et al., 2008) that may be responsible for increases in TOC and TN. The low mineralization rate of DL may have also caused the increase in TOC over the nine years, as compared to the baseline in 2002 (33g kg\(^{-1}\) soil), with TOC increases of 60, 30 and 10% in soils receiving DL\(_2\), DL\(_1\), and PS\(_2\), respectively. TOC in control soil did not increase, in comparison to the baseline, despite crop straw additions. This indicates that straw input is not sufficient to promote soil organic matter increase, probably because it is rapidly mineralized by soil microorganisms. However, in fertilized plots, the gain in biomass could be higher and the addition of crop straw could also influence the increases in TOC, due to increases in plant dry matter and nutrient content (LOURENZI et al., 2014).

Table 1 - Total organic carbon (TOC), total nitrogen (TN) and pH in unfertilized soil (C) or fertilized with pig slurry (PS\(_1\) and PS\(_2\)), deep-litter (DL\(_1\) and DL\(_2\)) and urea (U\(_1\) and U\(_2\)) at 104 and 209kg N ha\(^{-1}\) year\(^{-1}\).

| Treatments | TOC g kg\(^{-1}\) | TN g kg\(^{-1}\) | pH |
|------------|-----------------|-----------------|----|
| C          | 32.7 c          | 0.94 b          | 5.4 a |
| U\(_1\)    | 31.5 c          | 0.86 b          | 4.8 b |
| U\(_2\)    | 32.2 c          | 0.91 b          | 4.8 b |
| PS\(_1\)   | 33.1 c          | 1.00 b          | 5.6 a |
| PS\(_2\)   | 36.2 c          | 1.20 b          | 5.1 b |
| DL\(_1\)   | 44.2 b          | 1.60 a          | 5.7 a |
| DL\(_2\)   | 51.7 a          | 1.70 a          | 6.1 a |

Means followed by the same letter in each column do not differ significantly according to the Scott-Knott test at 5%.
Although the PS have a high concentration of N, this treatment did not increase the levels of this nutrient in the soil and this may be associated to its rapid immobilization in plants tissues and microorganisms, or to losses due to N volatilization (N₂O; NH₃) or leaching (NO₃⁻) (PLAZA et al., 2007). These processes happen rapidly since most of the PS nitrogen is in the ammoniacal N form (NH₃; NH₄⁺) and was placed in the soil surface, factor that increases losses over ten times (GIACOMINI et al., 2013). The soil fertilized with U also exhibited N losses, which can be due to rapid volatilization of this fertilizer.

The MBC showed no differences among control soil, DL₁, PS₁ and PS₂ (Figure 2a), but soils receiving DL₂ showed a 43% increase, while in soils receiving U₁ and U₂ MBC decreased by 13% and 32%, respectively, as compared to the control. PLAZA et al. (2007) showed that, MBC of soils PS-fertilized showed higher values (increments ranging from 30 to 50%) for several days after application, with gradual decreases

![Figure 2](image-url)
towards values close to control treatments, in periods ranging from 45 to 60 days. In this paper, the first phase of biomass growth is linked to degradation of manure labile compounds suggesting that, successive additions of liquid manure temporarily stimulate soil microorganisms. In soil receiving DL-, MBC increase is probably due to its higher nutrient and organic carbon content. Substrates with high organic content promote a more dynamic ecological succession, since they are attacked by organisms capable of degrading high-molecular weight molecules, while less specialized organisms can grow on the partially degraded substrates (GESSNER et al., 2010). MBC decreases in soils fertilized with urea may be due to the lack of C and micronutrients, which limit microbial growth and reproduction (GIACOMETTI et al., 2013). Decreases in MBC in soils with applications of U were associated with the adverse effects of drastic pH changes caused by this fertilizer (Table 1), as discussed above. According to a meta-analysis of MBC in soils from different biomes in Brazil, the values observed in this study are close to those reported for agricultural soils in the Atlantic Forest biome (491 to 591µg C g\(^{-1}\) dry soil), except for the treatments with U, which remained below the average values observed in the region (KASCHUK et al., 2010).

In general, soils receiving organic fertilizer had enhanced metabolic microbial activity, estimated as CO\(_2\) production (Figure 2b and 2c); with DL plots showing higher values than PS soils. The largest increases in soil microbial activity after application of pig manure occur in the first 30 days after manure application, and decrease to values similar to those in soils without addition of organic material (GIACOMINI et al., 2008). On the other hand, decreases in CO\(_2\) flow in U-amended soils (Figure 2d) can be explained by the lower carbon input associated to it. GIACOMETTI et al. (2013) showed that soils receiving only urea had lower CO\(_2\) production (reductions about 20%) when compared to those receiving organic amendments. Furthermore, soil pH decrease as caused by urea fertilization in the present research, inhibits growth of several groups of organisms. Season was other important factor for BR responses. From April to September, the coldest season, there were no differences in BR among plots amended with PS, DL and U. Low temperatures may have limited soil microbial activity, since temperature and soil moisture are primary factors controlling seasonal variation of CO\(_2\) production (GIACOMETTI et al., 2013).

Soils amended with PS\(_1\), DL\(_1\), DL\(_2\) and U\(_1\) showed the greatest \(\dot{q}\)CO\(_2\) values (0.45 to 0.52mg C-CO\(_2\) mg\(^{-1}\) MBC h\(^{-1}\)), and these treatments were higher than C, PS\(_2\) and U\(_2\) with values ranging from 0.32 to 0.35mg C-CO\(_2\) mg\(^{-1}\) MBC h\(^{-1}\). This can indicates stress in microbial communities, affected by changes in the community structure, substrate release or immobilization, and by soil management (WARDLE & GHANI, 1995; GIACOMETTI et al., 2013). Higher values of \(\dot{q}\)CO\(_2\) are found in microbial communities submitted to adverse conditions, in which microorganisms spend more energy (i.e., carbon compounds) for their maintenance in detriment of population growth (ANDERSON & DOMSCH, 2010). In the case of U-amended soils, the stress may be related to acidic pH and/or urea’s low C/N ratio, causing microorganisms to enter a high level of competition for carbon. On the other hand, successive fertilizations with PS and DL create a situation in which there may be high competition for available nutrients and constant changes in the community composition, which is based on substratum availability.

The fertilization treatments had no effect on enzyme activity, and the values were: β-glucosidase: 217.2µg PNP g\(^{-1}\) dry soil h\(^{-1}\); phosphatase: 277.3µg PNP g\(^{-1}\) dry soil h\(^{-1}\); aryl-sulphatase: 317.8µg PNP g\(^{-1}\) dry soil h\(^{-1}\); and FDA hyrolysis: 85.9µg PNP g\(^{-1}\) dry soil h\(^{-1}\). Some researches reported increases in activity of several enzymes in soils after a single application of swine manure (PLAZA et al., 2007; GIACOMETTI et al., 2013), but that may vary, for example, MARSCHNER et al. (2003) found no differences in enzymatic activity of soils with organic and mineral amendments in a long-term experiment and stated that enzyme activity measured may be masked, since the assays used indicate the potential activity, but not the actual enzyme activity in the soil.

There were changes in enzyme activity due to sampling time. In Oct. 2011 the activities of β-glucosidase, phosphatase, sulphatase, and FDA were 2.0; 3.8 and 3.6 times higher, respectively, than in Oct. 2010. This result, and the high correlation between enzyme activities and microbial biomass (r=0.67, P<0.01), suggest that enzyme activity increase is caused by the growth of microbial populations or by a shift in the microbial communities foraging different substrates (GESSNER et al., 2010). The activity of β-glucosidase and acid phosphatase is within the levels reported for these enzymes in other no-tillage soil systems in Southern Brazil. However, arylsulphatase activity was at least 10 times greater than those previously assessed (BALOTA et al., 2011). High arylsulphatase activity levels indicate increases in fungus biomass, since
this enzyme is related to ergosterol, a component of fungal cell membranes. No-tillage system can create favorable growth conditions for fungi, due to fewer disturbances. Thus, it is likely that the high activity of arylsulphatase recorded in this experiment is possibly associated with the soil management system adopted in the study area. Other studies in the same area have shown increases in soil carbon and significant changes in physical attributes due to application of swine manure (COMIN et al., 2013), associated with nutrients accumulation, such as phosphorus (GUARDINI et al., 2012). The present research has shown changes in soil microbial biomass and activity due to different forms of N fertilizers, which may impact soil quality as a whole.

CONCLUSION

Long term swine manure applications, in a no-tillage system, increase microbial activity and soil organic matter, while urea applications have negative impacts on these parameters. Swine manure amendments are useful for nutrient crop supply and integrated in a no-tillage system, significantly improve soil quality. Swine manure in deep litter form, is more efficient than slurry or urea in promotion of soil microbiota and carbon storage.

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