Review

Mining waste contaminated lands: an uphill battle for improving crop productivity

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Abstract: Mining drastically alters the physico-chemical and biological environment of the landscape. Low organic matter content, unfavourable pH, low water holding capacity, salinity, coarse texture, compaction, siltation of water bodies due to wash off of mineral overburden dumps, inadequate supply of plant nutrients, accelerated erosion, acid generating materials, and mobilization of contaminated sediments into the aquatic environment are the principal constraints experienced in mining contaminated sites. A variety of approaches have been considered for reclaiming mine wastes including direct revegetation of amended waste materials, topsoiling, and the use of capillary barriers. The simplest technology to improve crop productivity is the addition of organic amendments. Biosolids and animal manure can support revegetation, but its rapid decomposition especially in the wet tropics, necessitates repeated applications. Recalcitrant materials such as “biochars”, which improve soil properties on a long term basis as well as promote soil carbon sequestration, hold enormous promise. An eco-friendly and cost-effective Microbe Assisted Phytoremediation system has been proposed to increase biological productivity and fertility of mine spoil dumps. Agroforestry practices may enhance the nutrient status of degraded mine spoil lands (facilitation). N-fixing trees are important in this respect. Metal tolerant ecotypes of grasses and calcium-loving plants help restore lead, zinc, and copper mine tailings and gypsum mine spoils, respectively. Overall, an integrated strategy of introduction of metal tolerant plants, genetic engineering for enhanced synthesis and exudation of natural chelators into the rhizosphere, improvement of rhizosphere, and integrated management including agroforestry will be appropriate for reclaiming mining contaminated lands.

Keywords: agroforestry, biochars, mining sites, organic amendments, phytoremediation, revegetation

Introduction

Mining generates considerable waste materials and tailings, which are deposited on the surface as mine-spoil dumps. Removal of fertile topsoil, formation of unstable slopes prone to sliding and erosion, and siltation of water bodies due to wash off of mineral overburden dumps are also major negative effects of mining. The metals released from mining, smelting, forging, and other sources would accumulate in the soil (Khan et al., 2009), altering its chemistry. Metal contamination is not restricted to the mining site only because considerable release of metals occurs through acid mine drainage and erosion of waste dumps and tailing deposits (Salomons, 1995). Land use conflicts owing to operations close to the dwellings and farmlands as well as disposal of mine wastes on land intended for other uses, air and noise pollution, siltation of rivers by leachate and runoff from waste dumps, and degradation of land are also commonly associated with mining (Banda, 1995). Mining conflicts related to discharges of suspended solids rich in mercury, and cyanide from the artisanal gold mining areas into rivers are a major concern in Africa and South America. In the Ecuadorian artisanal gold mine environment, for instance it has been noticed that, when the river overflows, the mercury reaches the downstream banana plantations and shrimp ponds (UNIDO, 2007). The hazards of surface and groundwater pollution increases significantly when the mine waste materials contain reactive sulphide minerals such as pyrite (Liao et al., 2007). Pyrite-bearing mine tailings disposed at neutral or slightly alkaline conditions also can weather within a relatively short period of time to produce extreme acidity and lead to acid mine drainage (Robb and Robinson, 1995). Acid mine drainage usually contains a high load of heavy metals, in addition to having a low pH, which poses a major risk to surrounding water and soil systems (Achterberg et al.,
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2003; Braungardt et al., 2003). Chemical problems associated with surface mining, such as acid generating materials, are thus significant (Darmody et al., 2002) and in mine spoils, the geomorphic system is in disequilibrium (Dutta and Agarwal, 2001). Unfavourable soil chemistry and poor structure also deprive soil microbe and plant growth (Pederson et al., 1988).

Although mining-contaminated lands constitute a relatively small proportion of the total extent of degraded lands in the world, the scale of mining is increasing and the impacts are generally more severe than most other kinds of disturbances (Walker and Willig, 1999). Surface mined areas and mine spoil dumps are also nutritionally deprived habitats characterized by infertile soils having extreme pH values, low cation exchange capacity, low water holding capacity, low nutrient availability, and poor organic matter status (Gonzalez-Sangregorio et al., 1991).

Given the growing food insecure populations of the world (835.2 million undernourished people in the developing world according to FAO 2010), it is important to raise agricultural productivity on all types of lands including the mine-contaminated ones. In this paper, I will address the issues related to mine contaminated sites from the perspective of sustaining agricultural productivity.

Mine-spoil reclamation: problems and approaches

Reclamation of mine dumps and abandoned mine lands (AML) is a complex multi-step process. The first step in transforming the mine contaminated lands into productive agricultural lands is restoring its ecological integrity (Sheoran and Sheoran, 2009; Juwarkar and Singh, 2010). Most AML sites and many active mining or re-mining sites, however, lack any true topsoil and it consists primarily of mine spoil or overburden whose properties can range from loose, coarse textured material with many rock fragments, to highly compacted clay material. Broadly two types of effects are plausible: excesses (supra-optimal levels of chemical elements including metal ions) and deficiencies (suboptimal concentrations of essential elements).

Mine contaminated soils thus represent a very harsh environment for crop production (e.g., phytotoxicity and high acid production potentials of waste materials, low fertility, and limited topsoil availability). The principal restoration options are, therefore, ameliorative (improving the physical and chemical nature of the site) and adaptive (careful selection of species, cultivars, or ecotypes), both to be used in juxtaposition with one another (Johnson et al., 1994).

Amelioration

Topsoiling

Although a variety of approaches has been employed for reclaiming acid mine wastes including direct revegetation (no soil cover) of amended waste materials, application of topsoil is often the most effective method. Experimental studies have shown that topsoiling improved the water holding capacity and nutrient status of the mine wastes (e.g., Trlica et al., 1995), and provided a source of propagules and soil microorganisms (Schuman and Power 1980). Bowen et al. (2002) assessed the long-term (after 24 years) effects of different topsoil replacement depths (0, 20, 40, and 60 cm) on plant community cover, production, and diversity in south-central Wyoming, USA. Plant species richness was highest (7.5) at the zero topsoil depth and lowest (5.6) at the 60 cm topsoil depth. Total canopy cover was greatest (average 26.7%) at 40 and 60 cm of topsoil and least (21.5%) at the zero topsoil depth. Merrill et al. (1998) noted that the productivity of soil reconstructed by topsoil-subsoil placement on sodic mine spoil, however, would be influenced by the subsoil characteristics. Redente and Sydnor (2005) evaluated long-term plant community development on study plots in which 60 cm of retorted oil shale was covered by various depths of topsoil. Data over 20 years showed that native species were as productive as introduced species on deeper topsoil depths, implying the need for thick topsoiling. Excavated sediment of ponds and tanks is an effective indigenous soil amendment practice in India. Pond silt, rich in organic material, has been used for preparation of a topsoil layer of about 30–50 cm over the mine waste and levelled pits in some case studies (Singh et al., 2000; Wong, 2003). Silt layer also increased the productivity of the land and helped ground water recharge.

Capillary barriers

Despite potential benefits of amending mine waste and/or topsoiling, problems may arise such as acidification (or reacidification) of surface layers (Boon, 1986), excessive plant uptake of trace elements (Paschke et al., 2000), and/or capillary rise of soluble salts (McFarland et al., 1994). Therefore, some researchers have investigated the use of capillary barriers between overlying topsoil and underlying wastes as a reclamation option to reduce capillary rise of salts and trace elements and direct contact of plant roots with untreated waste materials. For instance, Molson et al. (2008) used covers with capillary barrier effects (CCBEs) for reducing acid mine drainage (AMD) from sulphidic mine tailings and found that capillary barrier covers significantly reduced sulphide oxidation and AMD. A CCBE
basically involves the placement of a relatively fine-grained soil, which acts as a water-retention layer, over a coarser capillary break material. Aubertin et al. (2009) showed that increasing the thickness of the cover may improve efficiency, but only up to a certain maximum beyond which the gain becomes minimal. One promising option is to combine different types of soil to create a layered CCBE.

**Phytoremediation**

Recently, the potential of bioremediation, particularly the role of higher terrestrial plants (phytoremediation) in reclamation of metal-polluted soils has been studied (Ghosh and Singh, 2005a), particularly for clean-up of diffusely polluted soils (Ginneken et al., 2007). Generally, decontamination of metal-contaminated soils requires the removal of toxic metals, as they cannot be degraded. Phytoremediation thus has emerged as a cost-effective, environment-friendly clean-up alternative. One of the directions in which research is currently evolving, is the use of oil-producing plant species, such as rape seed (Brassica napus) for phytoextraction purposes (Ginneken et al., 2007). Phytoremediation of metal contaminated soils thus is a ‘win-win’ situation: the biomass produced could be economically valorised in the form of bioenergy, e.g., Brassica spp. grown on metal contaminated sites can yield biodiesel, besides having the potential to accumulate high levels of heavy metals including Cd, Cr, Cu, Ni, Pb and Zn under certain conditions (Ebbs et al., 1997 and many others). A variety of factors, such as climatic conditions, soil properties, and site hydro-geology, however, may impact its efficiency (Lasat, 2000).

Short rotation coppice crops (SRC) consisting of fast growing trees such as willow (Salix spp.), poplar (Populus spp.), or black locust (Robinia pseudoacacia) are also promising as they can be used for bioenergy production and C sequestration (Quinkenstein et al., 2011), apart from phytoremediation. Experiments on reclamation sites have reported growth rates between 1 to 6 Mg/ha/yr for poplar and willow (Bungart and Hüttl, 2004; Grünewald et al., 2007). For black locust plantations established on reclamation sites in the mining district of Lower Lusatia, Germany, average aboveground biomass production ranged from 0.04 to 9.5 Mg/ha/yr for 1 to 14 years of growth (Quinkenstein et al., 2011).

Juwarkar and Singh (2010) suggested an eco-friendly and cost-effective *microbe-assisted phytoremediation (MAP)* approach for restoring zinc mine spoil dumps. This approach involves isolation and inoculation of site-specific specialised nitrogen-fixing strains of *Brdadyrhzobium* and *Azotobacter*, nutrient mobilising vesicular arbuscular mycorrhizal spores of *Glomus* and *Gigaspora* sp., selection of suitable plant species (preferably multispecies), and the use of organic amendments. This approach restored the productivity, fertility, and stability of zinc mine spoil leading to the development of sustainable ecosystems (Juwarkar and Singh 2010). Other beneficial microbes, which accumulates heavy metals, and decrease crop uptake include *Piriformospora indica*, a root-colonizing endophytic fungus (Oelmüller et al., 2009). Rhizobacteria, besides their role in metal detoxification/removal, also promote plant growth through production of growth promoting substances and siderophores (Khan et al., 2009). Choice of appropriate plant varieties, which are tolerant to the specific metals, is another design criterion for reclaiming mine overburdens. Among the tree species evaluated, *Eucalyptus tereticornis*, *Acacia auriculiformis*, and *Casuarina equisetifolia* were the most suitable for modification of spoil characteristics during the revegetation process (Dutta and Agrawal, 2002). Juwarkar et al. (2009) reported that tree species such as *Tectona grandis*, *Senna siamea*, *Dalbergia sissoo*, *Dendrocalamus strictus* and *Acacia nilotica* generated large biomass and soil organic matter, implying the need for site-specific selection of the tree and crop components.

**Mixed species stands and agroforestry**

In recognition of the role of trees to improve soil fertility (Nair et al., 2010), agroforestry systems (growing trees and crops in an integrated manner) are believed to have a great potential to reclaim the mine contaminated sites. This conjecture is based on the notion that tree incorporation would result in greater export of pollutants, improve site fertility, and render the sites productive. Since nutrient availability especially nitrogen commonly limits site productivity of mine spoils, the development of systems with nitrogen-fixing species—the so-called ‘fertilizer trees’ and cover crops—are important. Leguminous cover crops (e.g., *Centrosema*, *Calapogonium*, and *Pueraria*) are particularly important in this respect. Kimaro and Salifu (2011) also screened several grass species to identify appropriate cover crops for nursing newly planted tree seedlings and reported better survival and growth for trees grown in association with cover crops. Kumar et al. (1998) reported that in intercropping trials with teak (*Tectona grandis* + *Leucaena leucocephala*), teak growth increased linearly as the proportion of *Leucaena* in the mixture increased. At 44 months after planting, teak in the 1:3 teak-*Leucaena* mixture was 45% taller and 71% larger in diameter at breast height than those in pure stands. Likewise, Parrotta (1999) found that at harvest age of 4 years, total aboveground biomass ranged from 63 Mg/ha/yr in the *Eucalyptus* monoculture to 124 Mg/ha/yr in the
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_Cassuarina+Leucaena_ mixture (50:50), Kaye et al. (2000) compared N₂-fixers and non-N₂-fixers and found 20 to 100% more soil C under N₂-fixers. Overall, species mixtures especially those involving fertilizer trees may be useful in the ecorestoration of contaminated mine sites, where the soils are low in nitrogen.

**Soil Enhancement Applications**

**Organic matter and synthetic ameliorants**

Many studies have illustrated numerous benefits of adding organic matter (OM), in addition to lime and fertilizer, to acidic mine wastes (Bellitto et al., 1999). Mine reclamation research and practices have also demonstrated that organic amendments such as biosolids can support revegetation of mine spoil materials (Stehouwer, 1997). Most surface-mining reclamation operations also stockpile the A horizon and then redistribute these materials over the tailings and overburden to offset the problems of low water-holding capacity and compaction. Constructing an A horizon from tailing materials, however, requires additions of organic matter. By extension, productivity can be increased by adding various natural amendments such as saw dust, wood residues, sewage sludge, animal manures, and organic carbon to soil, which stimulates the microbial activity and augments nutrient (N, P) availability.

Synthetic and natural zeolites have been used as chelators for rapid mobility and uptake of metals from contaminated soils by plants (Prasad and Freitas, 1999). Use of synthetic chelators significantly increased Pb and Cd uptake and translocation from roots to shoots facilitating phytoextraction of the metals from low grade ores. Cross-linked polyacrylates, hydrogels, to metal-contaminated soils are used extensively (Prasad and Freitas, 1999) to increase the nutrient efficiency and alleviate the detrimental effects of the heavy metals (Prasad and Hagemeyer, 1999).

Hyper-accumulation can be induced in some plant species by soil amendment using EDTA on an insoluble target metal complex such as lead ore, rendering insoluble elements soluble (Anderson et al., 1998). Contrastingly, synthetic cross linked polyacrylates (hydrogels) have protected plant roots from heavy metal toxicity and prevented the entry of metals into roots (Prasad and Hagemeyer, 1999). However, large scale application of such synthetics may not be cost effective.

Organic matter inputs have the potential to improve the properties of mine tailings and spoils by increasing water-holding capacity, cation exchange capacity, buffering capacity, and by promoting soil structure and reducing bulk density (Smith et al., 1987). Incorporation of organic matter significantly increased aboveground biomass, with mushroom compost being more effective than biosolids (Redente and Sydnor, 2005). Various sources of organic resources and by-products are used as mine-spoil amendments. Recent research on these aspects is summarised below:

- Surface applications of municipal sewage sludge (Oyler 1988), fly ash (Moffat et al., 2001) and press mud (Juwarkar et al., 1992) were successful in promoting plant growth.
- Mine tailings amended with yardwaste compost (i.e., the end product of decomposing leaves and grass clippings) showed greater porosity, water-holding capacity (WHC), and saturated hydraulic conductivity of soil (K_{sat}) and lower mechanical resistance, and bulk density than un-amended tailings (Stolt et al., 2001).
- Because of the dominance of silt-size particles, fly ash may often be substituted for topsoil in surface mine lands, thereby enhancing physical conditions of soil, especially WHC (Adriano and Weber, 2001). Certain fly ashes are able to provide essential nutrients for plant nutrition.
- ‘Wastes’ from mining and mineral-related industries are useful for low-input agriculture. Examples include: waste from incomplete calcining in lime operations, calcium carbonate wastes from cement and other industries using CaCO₃, wastes from ‘black granite’ operations, phosphate mining / processing, steel production (e.g., basic slag, and calcium silicate slag), coal burning operations (fly ash, bottom ash, the by-products of fluidized bed combustion), and materials from flue gas desulphurization scrubbers are rich in micronutrients, with potential for application to the crop fields (van Straaten, 2002).
- Mine tailings containing biotite has the potential to be used as a slow-release K fertilizer and has been evaluated on pasture lands in Norway (Bakken et al., 2000). Likewise, calcium silicate slags increased sugarcane yield, specifically on low Si soils (Anderson et al., 1991).
- Phosphogypsum as a soil amendment on sodic soils and for groundnut fertilization has been tested (van Straaten, 2002). Pyrites and pyritic mill tailings with low to heavy metal contents have also been tested as inexpensive Fe-sources for sodic and Fe-deficient soils.
- Particle size and shape, alkalinity, and availability of several micronutrients in coal combustion by-products (CCBs) have been used to amend soil texture for increased water infiltration and acidity, and to supply some of the nutrient needs of the agricultural soils (Chugh et al., 2000). Alfalfa yields increased...
significantly by application of the CCB to the soil compared to the untreated control.

- Magnesium-containing fluidized bed combustion by-products have proved to be effective liming materials with a high effectiveness to ameliorate subsoil Al phytotoxicity (Stehouwer et al., 1999).

- Use of effluent treatment plant sludge, as an organic amendment, biofertilizers, and mycorrhizal fungi along with suitable plant species improved the physico-chemical properties of coal mine spoil (Juwarkar and Jambhulkar 2008).

**Charcoal and biochar application**

Considering the fact that soil organic matter in degraded land is very low, the simplest technology to improve soil productivity and stabilize crop yield is the addition of organic amendments, as outlined in the preceding section. The main limitation of organic matter addition, especially in the wet tropical condition, however, is its rapid decomposition, necessitating repeated additions during every planting season, which is impractical in view of the difficulty to source enough organic manures.

Some workers have therefore evaluated recalcitrant organic materials such as “biochars” for their ability to improve soil properties, carbon sequestration (e.g. Glaser et al., 2002; Lehman et al., 2003; Liang et al., 2006), and to increase crop yields (Yamato et al., 2006; Chan et al., 2008). Islami et al. (2011) tested the hypothesis that the beneficial effects of biochar as organic amendments in cassava based cropping system would last longer compared to that of the conventional organic manure such as farm yard manure and would promote soil carbon sequestration. The beneficial effects of biochar on soil properties also have been reported by many and includes chemical (Yamato et al., 2006), physical (Chan et al., 2008), and biological changes in the soil (Rondon et al., 2007). By extension, incorporation of biochar in mine contaminated sites may improve its nutrient retention power and productivity.

**Metal tolerant cultivars**

The use of metal tolerant ecotypes is a proven reclamation technology for lead, zinc, and copper mine tailings (Tordoff et al., 2000). Metal tolerance is a genetically heritable character, and some cultivars too have been bred incorporating this trait (e.g., *Festuca rubra* cv. “Merlin”) (Johnson et al., 1994). Direct seeding of tolerant cultivars is a promising area of further development. Results of some long term trials for exploiting biodiversity for dealing with difficult man-made substrates are available (e.g., Nicks and Chambers, 1995; Ginocchio, 1998).

Introduction of metal tolerant wild plants to metalliferous soils, genetic engineering of plants for enhanced synthesis and exudation of natural chelators into the rhizosphere, improvement of the rhizosphere with the help of mycorrhiza and integrated management of the metalliferous ecosystem following the principles of phytoremediation are important. The efficiency of phytoremediation can be enhanced by the judicious and careful application of appropriate heavy-metal tolerant, plant growth promoting rhizobacteria including symbiotic nitrogen-fixing organisms.

**Conclusions**

Future challenges in crop productivity of the mining and mineral industries include the increasing scale of operations with large mining companies seeking to exploit large reserves in more remote wilderness environments, greater innovation in new technologies such as the *in situ* extraction of metals through leaching, the increasing need to regulate and develop environmental management in the artisanal and small mining sector, and the imperative to incorporate policies of sustainable development as far as possible. Most of the new mining initiatives currently are in developing countries, and this will extend to mining ore deposits in more remote and fragile ecosystems. The time has arrived for a rethinking on the way mine contaminated site development programs are planned and implemented around the world.

We need to encourage the “remarriage of trees, tolerant crops and microbes” on these landscapes and exploit the time-tested benefits of such practices to address some of the major threats of increasing extent of mine spoils and contamination. If we are to meet society’s needs and aspirations, we must find novel ways of utilizing mine contaminated lands.

While it is creditable that considerable progress has been achieved during the past three decades in transforming mine-contaminated lands into agriculturally productive sites, several knowledge gaps exist even in areas that have received research attention in the past. There are also several potentially promising areas that have not yet been explored. For example, substantial efforts are needed to domesticate metal tolerant species and breed new cultivars with higher yield potential. In our obsession with “grain crops” in modern agriculture, we have ignored the tree component which has considerable potential for “phytoremediation”.

The exploitation of these species, and the agroforestry practices involving their use, has wide implications in food security and environmental protection, as well as conservation and use of genetic resources.
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References

Achterberg, E.P., Herzl, V.M.C., Braungardt C.B. and Millward G.E. 2003. Metal behaviour in an estuary polluted by acid mine drainage: The role of particulate matter. Environmental Pollution 121:283–292.

Adriano, D.C. and Weber, J.T. 2001. Influence of fly ash on soil physical properties and turfgrass establishment. Journal of Environmental Quality 30:596–601.

Anderson, C.W.N., Brooks, R.R., Stewart, R.B. and Simcock R. 1998. Harvesting a crop of gold in plants. Nature 395:553–554.

Anderson, D.L., Snyder, G.H. and Martin, F.G. 1991. Multi-year response of sugarcane to calcium silicate slag on Everglade histosols. Agronomy Journal 83:870–874.

Aubertin, M., CIFUENTES, E., Apithy, S.A., Bussière B., Molson J. and Chauquisa R.P. 2009. Analyses of water diversion along inclined covers with capillary barrier effects. Canadian Geotechnical Journal 46: 1146–1164.

Bakken, A.K., Gautneb, H., Sveistrup, T. and Myhr, K. 2000. Crushed rocks and mine tailings applied as K fertilizers on grassland. Nutrient Cycle in Agroecosystem 56:53–57.

Banda, P.M. 1995. Environmental Management in the Mineral Sector in Zambia. Proceedings of 1995 international conference on industrial minerals: Investment Opportunities in Southern Africa, Mambwe. S H, Simukanga, S, Sikazwe, O N and Kamona, F (eds). 7–9 June 1995, Pomadzi Hotel Lusaka.

Bellitto, M.W., Williams, H.T. and Ward, J.N. 1999. Application of ameliorative and adaptive approaches to revegetation of historic high altitude mining wastes. In: Bengson S.A. and Bland D.M. (ed.). Proc. 16th Ann. Meeting of the Am. Soc. for Surface Mining and Reclamation. Am. Soc. for Surface Mining and Reclamation, Scottsdale, AZ. pp 165–174.

Boon, D.Y. 1986. Identification and handling of acid materials. In M.A. Schuster and R.H. Zuck (ed.) Proc. High Altitude Revegetation Workshop no. 7. Colorado State Univ. Info. Ser. 58. Colorado State Univ., Fort Collins, pp 49–65.

Bowen, C.K., Olson, R.A., Schuman, G.E. and Ingram L.J. 2002. Long-term plant community responses to topsoil replacement depth on reclaimed mined land. Paper was presented at the National Meeting of the American Society of Mining and Reclamation, Lexington KY, 9–13 June 2002. Published by ASMR, 3134 Montavesta Rd., Lexington, KY 40502, pp 130–140.

Braungardt, C.B., Achterberg, E.P., Elbaz-Pouliquet, F. and Morley, N.H. 2003. Metal geochemistry in a mine-polluted estuarine system in Spain. Applied Geochemistry 18:1757–1771.

Bungart, R. and Hüttl, R.F. 2004. Growth dynamics and biomass accumulation of 8-year-old hybrid poplar clones in a short-rotation plantation on a clayey-sandy mining substrate with respect to plant nutrition and water budget. European Journal of Forestry Research 123: 105–115.

Chan, K.Y., Van Zwieten, B.L., Meszaros, I., Downie, D. and Joseph, S. 2008. Using poultry litter biochars as soil amendments. Australian Journal of Soil Research 46: 437–444.

Chugh, Y.P., Deb, D. and Raju, C.B. 2000. Physical and engineering properties of coal combustion by-products. Proceedings of the use and disposal of coal combustion by-products at coal mines: A technical interactive forum. The National Energy Technology Laboratory, Morgantown, West Virginia, 10–13 April 2000. Vories K.C. and Throgmorton D. (ed.). National Energy Technology Laboratory, U.S.A. pp 26–33.

Darmody, R.G., Dunker, R.E. and Barnhisel, R.I. 2002. Reclamation of prime agricultural lands after coal surface mining: the Midwestern experience. National Meeting of the American Society of Mining and Reclamation, Lexington KY, June 9–13, 2002. Published by ASMR, 3134 Montavesta Rd., Lexington, KY 40502, pp 900–915.

Dutta, R.K. and Agrawal, M. 2001. Litterfall, litter decomposition and nutrient release in five exotic plant species planted on coal mine spoils. Pedobiologia 45: 298–312.

Dutta, R.K. and Agrawal, M. 2002. Effect of tree plantations on the soil characteristics and microbial activity of coal mine spoil land. Tropical Ecology 43: 315–324.

Ebbes S.D., Lasat M.M., Brady D.J., Cornish R., Gordon R. and Kochian L.V. 1997. Phytoextraction of cadmium and zinc from a contaminated soil. Journal of Environmental Quality 26:1424–1430.

FAO. 2010. The State of Food Insecurity in the World Addressing food insecurity in protracted crises. Food and Agriculture Organization of the United Nations, Rome, 58p.

Ghosh, M. and Singh .S.P. 2005. A review on phytoremediation of heavy metals and utilization of its byproducts. Applied Ecology and Environmental Research 3: 1–18.

Ginneken, L.V., Meers, E., Guinnocchio, R. 1998. Chile: Restoration challenges. In: A.G. Lacombat, M. Ginocchio, R. (eds). Proceedings of the use and disposal of coal combustion by-products at coal mines: A technical interactive forum. The National Energy Technology Laboratory, Morgantown, West Virginia, 10–13 April 2000. Vories K.C. and Throgmorton D. (ed.). National Energy Technology Laboratory, U.S.A. pp 26–33.

Glaser B., Lehmann J. and Zech W. 2002. Ameliorating physical and chemical properties of highly weathered soils in the tropics with charcoal: A review. Biology and Fertility of Soils 35: 219–230.

Gonzalez-Sangregorio, M.V., Trasar-Cepeda M.C., Leiros M.C., Gil-Sotres F. and Guinian-Ojea F. 1991. Early stages of lignite mine soil genesis: Changes in biochemical properties. Soil Biology and Biochemistry 23: 589–595.

Grünewald, H., Brandt, B.K.V., Schneider, B.U., Bens, O., Kendzia, G. and Hüttl, R.F. 2007 Agroforestry systems for the production of woody biomass for energy transformation purposes. Ecological Engineering 29: 319–328.
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Isalm, T., Guritno, B., Basuki, N. and Suryanto, A. 2011. Biochar for sustaining productivity of cassava based cropping systems in the degraded lands of East Java, Indonesia. Journal of Tropical Agriculture 49: 40–46.

Johnson, M.S., Cooke, J.A. and Stevenson, J.K. 1994. Revegetation of metalliferous wastes and land after metal mining. In: Mining and its Environmental Impact. Hester R.E. and Harrison R.M. (eds). Issues in Science and Technology, Royal Society of Chemistry, Letchworth, England, pp. 31–48.

Juwarkar, A.A. and Jambhulkar, H.P. 2008. Phytoremediation of coal mine spoil dump through integrated biotechnological approach. Bioresources Technology 99: 4732–4741.

Juwarkar, A.A. and Singh, S.K. 2010. Microbe-assisted phytoremediation approach for ecological restoration of zinc mine spoil dump. International Journal of Environmental Pollution, 43:236–250.

Juwarkar, A.A., Yadav, S.K., Thawale, P.R., Kumar P., Singh S.K. and Chakrabarti T. 2009. Developmental strategies for sustainable ecosystem on mine spoil dumps: a case of study. Environmental Monitoring and Assessment 157:471–483.

Juwarkar, A.S., Juwarkar, A., Pande, V.S. and Bal, I.S. 1992. Restoration of manganese mine spoil productivity through pressmud utilization. In R.K. Singhal, A.K. Malhotra and J.L. Collins (eds). Environmental Issue and Management of Waste Energy and Production. Brookfield: Balkema, pp. 827–830.

Kaye, J. P., Resh, S. C., Kaye, M. W. and Chimner, R. A. 2000. Nutrient and carbon dynamics in a replacement series of Eucalyptus and Alibizia trees. Ecology 81:3267–3273.

Khan, M.S., Zaidei, A., Wani, P.A. and Oves, M. 2009. Role of plant growth promoting rhizobacteria in the remediation of metal contaminated soils. Environmental Chemistry Letters 7:1–19.

Kimaro, A.A. and Salifu, K.F. 2011. The effect of cover crop species on growth and yield response of tree seedlings to fertiliser and soil moisture on reclaimed sites. In: Mine Closure 2011. Fourie A.B., Tibbett M., and Beersing A. (eds) Australian Centre for Geomechanics, Perth, pp 37–46.

Kumar, B.M., Kumar, S.S. and Fisher, R.F. 1998. Intercropping teak with Leucaena increases tree growth and modifies soil characteristics. Agroforestry System 42: 81–89.

Lasat, M.M. 2000. Phytoextraction of metals from contaminated soil: a review of plant/soil/metalinteraction and assessment of pertinent agronomic issues. Journal of Hazardous Substances Research 2: 5-1 to 5-25.

Lehman, J., da Silva Jr., J.P., Steiner, C., Nehls, T., Zech, W. and Glaser, B. 2003. Nutrient availability and leaching in an archaeological Anthrosol and a Ferralsol of the Central Amazon basin: fertilizer, manure and charcoal amendments. Plant Soil 249: 343–357.

Liang, B., Lehmann, J., Kinyangi, D., Grossman, J., O’Neill, B., Skjemstad, J.O., Thies, J., Luizao, F., Peterson, J. and Neves, E.G. 2006. Black carbon increases cation exchange capacity in soils. Soil Science Society of America Journal 70: 1719–1730.

Liao, B., Huang, L.N., Ye, Z.H., Lan, C.Y. and Shu, W.S. 2007. Cut-off net acid generation pH in predicting acid-forming potential in mine spoils. Journal of Environmental Quality 36:887–891.

McFarland, M.L., Ueckert, D.N., Hons, F.M. and Hartmann S. 1994. Selective-placement burial of drilling-fluids—Effects on soil properties, buffalograss and forwiring salbush after 4 years. Journal of Rangeland Management 47:475–480.

Moffat, A. J., Armstrong, A. T. and Ockleston, J. 2001. The optimization of sewage sludge and effluent disposal on energy crops of short rotation hybrid poplar. Biomass Bioenergy 20:161–169.

Molson, J., Aubertin, M., Bussière, B. and Benzaazoua, M. 2008. Geochemical transport modelling of drainage from experimental mine tailings cells covered by capillary barriers. Applied Geochemistry 23: 1–24.

Nair, P.K.R., Nair, V.D. Kumar, B.M. and Showalter, J.M. 2010. Carbon sequestration in agroforestry systems. Advance Agronomy 108: 237–307.

Nicks, L.J., and Chambers, M.F. 1995. Farming for metals? Mining Environmental Management 3: 15–18.

Oyler, J. 1988. Revegetation of metals contaminated site near a zinc smelter using sludge/fly ash amendments and herbaceous species. Trace Substances in Environmental Health 22: 306–320.

Parrotta, J.A. 1999. Productivity, nutrient cycling, and succession in single- and mixed-species plantations of Casuarina equisetifolia, Eucalyptus robusta, and Leucaena leucocephala in Puerto Rico. Forest Ecology and Management 124: 45–77.

Paschke, M.W., Redente, E.F. and Levy, D.B. 2000. Zinc toxicity thresholds for important reclamation grass species of the western United States. Environmental Toxicology and Chemistry 19:2751–2756.

Pederson, T.A., Rogowski, A.S. and Pennock, R. 1988. Physical characteristics of some mine spoils. Soil Science Society of America Journal, 44: 131–140.

Prasad, M.N.V. and Freitas, M. De.O.H. 1999. Feasible biotechnological and bioremediation strategies for serpentine soils and mine spoils. EJB Electronic J. Biotechnol. 2.1, April 15, 1999. Available at http://www.ejb.org/content/vol2/issue1/full/5 (accessed 15 Jan 2012).

Prasad, M.N.V. and Hagemeyer, J. (eds) 1999. Heavy metal stress in plants - From molecules to ecosystems. Springer- Verlag, Heidelberg, Berlin, New York, 401p.

Quinckenstein, A., Böhm, C., Matos, E., Freese D., and Hüttl, R.F. 2011. Assessing the carbon sequestration in short rotation coppices of Robinia pseudoacacia L. on marginal sites in northeast Germany. In: Carbon sequestration potential of agroforestry systems: opportunities and challenges. Kumar, B.M. and Nair, P.K.R. (eds). Springer Science, The Netherlands, pp 201–216.

Redente, E.F. and Sydnor, R.S. 2005. Long-term plant community development on topsoil treatments overlying a phytotoxic growth medium. Paper was presented at the 2005 National Meeting of the American Society of Mining and Reclamation, Breckenridge CO, June 19–23 2005. Published by ASMR, 3134 Montavesta Rd., Lexington, KY 40502, pp 963–975.

Oelmüller, R., Sherameti, I., Tripathi, S. and Varma, A. 2009. Piriformospora indica, a cultivable root
endophyte with multiple biotechnological applications. *Symbiosis* 49:1–17.

Robb, G.A. and Robinson, J.D.F. 1995. Acid drainage from mines. *Geography Journal* 161:47–54.

Rondon, M.A., Lehmann, J., Ramirez, J. and Hurtado, M. 2007. Biological nitrogen fixation by common beans (*Phaseolus vulgaris* L.) increases with biochar additions. *Biology and Fertility of Soils* 43: 699–708.

Salomons, W. 1995. Environmental impact of metals derived from mining activities-processes, predictions, prevention. *Journal of Geochemical Exploration* 52: 5–23.

Schuman, G.E. and Power, J.F. 1980. Plant growth as affected by topsoil depth and quality on mined lands. In: *Adequate Reclamation of Mined Lands?* Symp. 26–27 Mar. 1980. Soil Conserv. Soc. of Am. and WRCC-21, Billings, MT. pp. 6–1 to 6–9.

Sheoran, V. and Sheoran, A.S. 2009. Reclamation of abandoned mine land. *Journal of Mining and Metallurgy* 45 A (1): 13–32.

Singh, A., Jha, A.K. and Singh, J.S. 2000. Effect of nutrient enrichment on native tropical trees planted on Singrauli Coalfields, India. *Restoration Ecology* 8: 80–86.

Smith, P.L., Redente, E.F. and Hooper, E. 1987. Soil organic matter. In Reclaiming Mine Soils and Overburden in the Western United States: Analytical Parameters and Procedures. R.D. Williams and G.E. Schuman (eds.). Soil Conservation Society of America, Akeny, IA, pp. 185–214.

Stehouwer, R.C. 1997. Amendment of acidic coal refuse with yard waste compost and alkaline materials: Effects on leachate quality and plant growth. *Compost Science and Utilization* 5:81–92.

Stehouwer, R.C., Dick, W.A. and Sutton, P. 1999. Acidic soil amendment with a magnesium-containing fluidized bed combustion by-product. *Agronomy Journal* 91:24–32.

Stolt, M.H., Baker, J.C., Simpson, T.W., Martens, D.C., McKenna, J.R. and Fulcher, J.R. 2001. Physical reconstruction of mine tailings after surface mining mineral sands from prime agricultural land. *Soil Science* 166: 29–37.

Tordoff, G.M., Baker, A.J.M. and Willis, A.J. 2000. Current approaches to the revegetation and reclamation of metalliferous mine wastes. *Chemosphere* 41: 219–228.

Trlica, M.J., Brown, L.F., Jackson, C.L. and Jones, J. 1995. Depth of soil over molybdenum tailing as it affects plant cover, production, and metal uptake. In W.R. Keammerer and W.G. Hassell (ed). *Proc. High Altitude Revegetation Workshop*, no. 11. Colorado State Univ. Info. Ser. 80. Colorado State Univ., Fort Collins, pp. 119–145.

UNIDO. 2007. The Artisanal Gold Mining: Case study of mercury and cyanide in Ecuador. Technical report: Based on the work of Mr. Patricio C. Velasquez L. United Nations Industrial Development Organization, Vienna, 69p.

van Straaten, P. 2002. *Rocks for Crops: Agrominerals of Sub-Saharan Africa*. ICRAF, Nairobi, Kenya, 338p.

Walker, L.R. and Willig, M.R. 1999. An introduction to terrestrial disturbances. In *Ecosystems of disturbed ground*. *Ecosystems of the World*. Walker L. (ed.). Elsevier, Amsterdam. pp. 1–16.

Wong, M.H. 2003. Ecological restoration of mine degraded soils, with emphasis on metal contaminated soils. *Chemosphere* 50: 775–780.

Yamato, M., Okimori, Y., Wibowo, I.F., Anshori, S. and Ogawa, M. 2006. Effects of the application of charred bark of *Acacia mangium* on the yield of maize, cowpea and peanut, and soil chemical properties in South Sumatra, Indonesia. *Journal of Soil Science and Plant Nutrition* 52: 489–495.