Determining spatio-temporal distribution of bee forage species of Al-Baha region based on ground inventorying supported with GIS applications and Remote Sensed Satellite Image analysis

Nuru Adgaba, Ahmed Alghamdi, Rachid Sammoud, Awraris Shenkute, Yilma Tadesse, Mahammad J. Ansari, Deepak Sharma, Colleen Hepburn

Engineer Bagshan Chair for Bee Research, Department of Plant Protection, College of Food and Agricultural Science, King Saud University, P O Box: 1460, Riyadh 11451, Saudi Arabia
College of Computer and Information Sciences, King Saud University, Saudi Arabia
Department of Zoology, Rhodes University, South Africa

Article history:
Received 26 March 2016
Revised 5 December 2016
Accepted 1 January 2017
Available online 24 January 2017

Keywords:
Bee forage
Plant inventory
GIS
Satellite image
Flowering period
Saudi Arabia

Abstract
In arid zones, the shortage of bee forage is critical and usually compels beekeepers to move their colonies in search of better forages. Identifying and mapping the spatiotemporal distribution of the bee forages over given area is important for better management of bee colonies. In this study honey bee plants in the target areas were inventoried following, ground inventory work supported with GIS applications. The study was conducted on 85 large plots of 50 × 50 m each. At each plot, data on species name, height, base diameter, crown height, crown diameter has been taken for each plant with their respective geographical positions. The data were stored, and processed using Trimble GPS supported with ArcGIS10 software program. The data were used to estimate the relative frequency, density, abundance and species diversity, species important value index and apicultural value of the species. In addition, Remotely Sensed Satellite Image of the area was obtained and processed using Hopfield Artificial Neural Network techniques. During the study, 182 species from 49 plant families were identified as bee forages of the target area. From the total number of species; shrubs, herbs and trees were accounting for 61%, 27.67%, and 11.53% respectively. Of which *Ziziphus spina-christi*, *Acacia tortilis*, *Acacia origina*, *Acacia asak*, *Lavandula dentata*, and *Hypoestes forskafoi* were the major nectar source plants of the area in their degree of importance. The average vegetation cover values of the study areas were low (<30%) with low Shannon’s species diversity indices (H’) of 0.5–1.52 for different sites. Based on the eco-climatological factors and the variations in their flowering period, these major bee forage species were found to form eight distinct spatiotemporal categories which allow beekeepers to migrate their colonies to exploit the resources at different seasons and place. The Remote Sensed Satellite Image analysis confirmed the spatial distribution of the bee forage resources as determined by the ground inventory work. An integrated approach, combining the ground inventory work with GIS and satellite image processing techniques could be an important tool for characterizing and mapping the available bee forage resources leading to their efficient and sustainable utilization.

1. Introduction

The Kingdom of Saudi Arabia extends over two million km² land area. Despite its arid climatic conditions, the country consists of diverse ecologies and floras in which more than two thousands of plant species have been recorded (Collenette, 1999; Chaudhary, 2001). As a result, beekeeping is practiced in many areas of the country and it is one of the important income generating activities for rural communities. Despite the presence of diverse number of bee plants, seasonal shortage of bee forage is critical. These conditions force many beekeepers to move their...
colonies from place to place (averagely 5 times/year) following the availability of bee forage in different ecologies and seasons (Nuru et al., 2014). This often led to overcrowding of large numbers of bee colonies in a few areas, causing intense competition for bee forage and the subsequent declines of productivity of colonies (Khanbash, 2001; Al-Ghamdi, 2007). However, the type of honey bee plants, their relative abundance, spatial distribution, phenology and their relative values for bees and honey production were not studied and documented.

Honey bee floral resources vary in ecological distribution and in periods of availability according to their flowering times (Chemas and Rico-Gray, 1991). It is therefore, very important to differentiate honey bee plants from the total plants, to map their spatial and temporal distributions and determine their relative values to bees and honey production for every geographical region. Such types of information are important to categorize bee forage plants as useful for colony population build-up and honey production (Partap, 1997).

To identify and inventory honey bee plants; different approaches such as palynological analysis of honey samples, direct observation of foraging bees and analysis of pollen loads and pollen stores in the nest have been used (Amoko, 1997; Hepburn and Jacot-Guillarmod, 1991; Hepburn and Radloff, 1995; Admasu, 2003; Abou-Shaara, 2015a,b). However, these methods only give a general idea as to which plant species the honey bee foragers collected pollen from, and not other co-existing plant species which may flower concurrently and from which resources are also readily available. In recent years, another approach is that of above ground vegetation biomass assessments have been used to estimate the bee forage potentiality of an area. The most common techniques are estimations from field measurements which involve representative sampling following the protocols such as by Kent and Coker (1992), Peichl and Arain (2006), Wulder et al. (2008), or direct measurements like those of Suganuma et al. (2006).

Despite their merits in degree of accuracy and applicability for specific tree species, ground inventory work is highly time and labor intensive and difficult to cover large areas. Peichl and Arain (2007) showed the presence of a strong correlation between spectral information recorded by Remote Sensing of Satellite Images with that of vegetation conditions. In addition Sahoo et al. (2005) used satellite-derived images showing temporal vegetation phenology (green-up and brown-down) to monitor vegetation changes. Several vegetation indices particularly, NDVI reflectance measurements like those of Suganuma et al. (2006). Gould (2000) and Rocchini et al. (2007) have used Remote Sensed Satellite Images to map vegetation cover down to species-level.

However, satellite images in their raw form are not suitable to understand and interpret in a meaningful way and to use them for specific objectives. Therefore, it requires a suitable image segmentation process, which is partitioning of an image into non-overlapping, meaningful and homogenous regions based on their pixel properties (Sammouda et al., 2013). Pixel clustering is one of the effective and efficient techniques used to create homogeneous image regions for image segmentation and its aim is to partition the desired clusters with high intra-class similarities (Felzenszwalb and Huttenlocher, 2004). In the satellite image segmentation, the application of the Hopfield Artificial Neural Network (HNN) clustering algorithm technique is reported to be useful to process the images in meaningful homogenous regions (Shi and Mallick, 2000).

Since different tree species become green-up as a result of flushing of leaves at different times, Remote Sensed Satellite Image processing of such changes, supported with GIS based ground inventory work, can be used to analyze the temporal and spatial distribution of bee forage species at different seasons. The aim of the current work was to determine the spatiotemporal distribution of honey bee forages of the target areas through detail ground inventory work using GIS applications and supported with Remote Sensed Satellite Image processing techniques.

2. Materials and methods

2.1. Study area

The study was conducted in Al-Baha region, Saudi Arabia (Fig. 1). The region is characterized by diverse physiographic conditions which vary from lowland plains to highland plateaus with altitudinal range of 300–2400 m above sea level. The topography of the region is mainly undulated hills, mountains with strips of fertile valleys. The average relative humidity of the region ranges from 52 to 67%. The rainfall amount is relatively low and varies from 229 to 581 mm/annum. The mean temperature is 22.9℃.

2.2. Ground inventory

The vegetation composition of the target areas was determined by taking 11 representative study sites (valleys). A total of 85 large sampling plots (50 × 50 m each) (total sampling area of 212,500 m²) were covered. The sampling plots were taken following Kent and Coker (1992), Flombaum and Sala (2007) and Suganuma et al. (2006) protocols. Unlike the previous protocols, we used large sampling plots, because trees and shrubs found very scattered in the study area. For each plot all the necessary informa-
tion like: plant species type, height, basal diameter, crown height, crown diameter were taken for each plant with its respective geographical position. The data were entered in Trimble® Juno® 3 series, GPS (USA). To facilitate the data analysis, a database was designed using the ArcGIS-10 software program and used to store, automate and process the entered data. The inventory of honey bee plant species of the target area was conducted by registering all the bee plant species after authentication of the plants as bee forage based on visitation of honey bees to the species for pollen or/and nectar. Voucher specimens were collected of plants for which the species names were unknown, these were sent to the Department of Botany, King Saud University, for identification.

The ground survey data were captured as layers in the database application. The list of entered attributes of each measured tree in each valley was used as vector information to analyze the vegetation condition of the study area. From these data, the relative frequency, relative density, relative abundance, species diversity indices, species important value indices (IVI) and apicultural value (AV) were calculated. The apicultural value of the species was calculated by multiplying the species important value indices with the average nectar sugar amount per plant. The average nectar sugar per plant was determined following Mallick (2000), Kim et al. (2011), Nuru et al. (2012) protocols with some modifications.

2.3. Satellite image processing

To support the ground inventory survey, Remote Sensed Satellite Images (1.0 m resolution) of the study area at different seasons of the year 2013 were obtained from the King Abdulaziz City for Science and Technology Space Research Institute. Processing (color segmentation) of the satellite images was done by assigning (labeling) each pixel with very close features (color, intensity and texture) into only one cluster. This method uses the pixel label to separate objects in the image based on the mean color value of the corresponding cluster in the original satellite image. Each cluster varies from the other depending on spectral information of the forest biomass images recorded by Remote Sensing. Segmentation of the satellite image color was done using the Hopfield Artificial Neural Network (HNN) clustering algorithm technique (Shi and Mallick, 2000). The segmentation was supported with Computer Aided Determination programs using TrueQuant Software application.

2.4. Phenology

Along the inventory of the honey bee forage species their flowering calendar: the onset of flowering, peaking, ending, and total duration of flowering has been recorded by continuous monitoring of the phenology of each species. The monitoring was done at least for two flowering seasons for each plant species.

3. Results and discussion

3.1. Ground inventory

During the survey, 182 plant species were identified as pollen or/and nectar sources for honey bees in the region. The bee forage species were distributed over 49 plant families of which the five families, Asteraceae, Leguminosae, Lamiaeae, Acanthaceae, and Malvaeae were accounted for the majority (35%) of the total bee forage species of the region. Of the total number of species recorded 111 (61%) were shrubs, 50 (27.67%) herbs and 21 (11.53%) trees. Despite there being less species diversity; trees and shrubs were the major sources of honey in the region, which could be due to their deep rooted nature and their adaptation to low precipitation. Moreover, trees and shrubs may not be affected equally by intermittent rainfall conditions of the area as for instance annual herbs. Of the total 182 species recorded as bee plants only six species (Ziziphus spina-christi, Acacia tortilis, Acacia origina, Acacia asak, Lavandula dentata, and Hypoestes forskoeli) serve as major sources of nectar in the region. This is in agreement with previous reports that mentioned only 1.6% of the world’s bee plants are the source of most of the world’s honeys (Crane, 1990). Z. spina-christi was reported as a major honey source plant of eastern Saudi Arabia (Taha, 2015). The importance of having lists of honey bee plants of an area in suitable apiary site selection and cultivating of potential plants for honey bee colonies has been well suggested (Abou-Shaara, 2015a,b).

Species such as Dipterygium glaucum, Plantago lanceolata, Acacia spp., Psidadia punctulata, Rumex spp. and many grass species are very good sources of pollen for honey bees in the region. In general, pollen source plants are very limited, both in terms of number of species, and their distribution and duration of availability. As a result, brood rearing and colony build-up are very critical to beekeeping in the region. Major reasons for the shortage of pollen could be limited rainfall to support the growth and flowering of annual bee forage species, and the absence of large-scale cultivation of annual crops that serve as additional sources of pollen. Shortage of pollen as a result of extended dry periods and recurrent droughts and subsequent declining of colony performances has been reported in Saudi Arabia (Omar et al., 2013).

The flowering periods of the bee forage plants of the study area were fairly distributed among the different seasons, a higher proportion of flowering takes place in spring (35%) followed by summer (24.4%). Similarly Taha (2015) has reported that high proportion of flowering of bee plants takes place during spring for eastern parts of Saudi Arabia. Some species such as Z. spina-christi, Acacia etbaica, A. johnwoodii and A. asak have multiple flowering periods. Moreover, plant species such as Z. spina-christi, A. tortilis and Acacia ehrengberiana have a wide range of ecological distribution that ranges from 200 to 1750 meters above sea level. As a result, the flowering periods of the same species vary between ecologies.

Generally, the Acacia species which are the dominant bee forage species in the areas are spatio-temporally structured in such a way that some Acacia species (A. johnwoodii and A. ehrengberiana) spatially co-exist but have different flowering seasons. While others such as A. tortilis and A. ehrengberiana co-exist and flower concurrently but vary with respect to the time of peak flowering within a season. Furthermore, other Acacia species (A. etbaica, A. johnwoodii and A. asak) are spatially separated in that A. asak has been observed to grow mainly in steep rocky escarpments, whereas A. johnwoodii grows mainly in the bottoms of valleys or on gently sloping ground with deep soil structure. A. etbaica grows on steep slopes or on gently sloping ground, but favoring the eastern aspects of landscape, and does not spatially overlap with other species which flower concurrently.

All of these phenomena may be considered as adaptations by the species to avoid competition for pollinators and minimize heterospecific pollen transfer among related species. Similarly, Stone et al. (1998) recognized similar mechanisms as major contributing factors to the structuring of the flowering of sympatric Acacia species in Africa. A mechanism for the minimization of competition for pollinators is believed to be an important force in the structuring of plant communities (Feinsinger et al., 1987). The spatial variation of the Acacia species reported to be partially attributed to avoiding completion for pollinators, (Rathcke, 1988). Besides the competition for pollinators; other environmental factors such as physiographic (altitude, soil, slope) and climatic conditions believed to play important role in spatiotemporal distribution of plant species.
Environmental factors, such as variations in the amount of water stored by plants (Borchert, 1994a,b), and changes in temperature (Williams-Linera, 1997) are regarded as important factors involved in phenological variations among some tropical plants. The schematic map of the temporal and spatial distribution of the major nectar and pollen source species of the region are depicted in Fig. 2. This enables beekeepers to harvest honey two times in the same season from the same species.

Generally, flowering periods are very short and intermittent in the region. However, deep rooted trees and shrubs such as Z. spina-christi and many of the Acacia spp. flower, more or less, following regular seasons. However, flowering periods of the majority of annuals, biennials and some perennials, are governed by the onset of rainfall and subsequent soil moisture conditions. As a result, flowering periods’ shift from year to year and place to place is highly expected. The effect of rainfall on growth, green up and onset of flowering of plants in different climatic regions have been well documented (Fox, 1990; Abd-ElGhani, 1997; Peñuelas et al., 2004). Moreover, the general spatiotemporal phenological shifts of many plant species as responses to changes in rainfall are also well known (Peñuelas et al., 2004).

Moreover, some plants like A. tortilis and A. ehrenbergiana were observed to flower in dry season, in leafless stages, and secrete considerable amounts of nectar from stored carbohydrates of the previous season (pers. obs. Nuru). However, when rainfall occurs, the plants shift their resource allocation from nectar secretion to vegetative growth (i.e. producing new leaf and vegetative buds) which may lead to declining of nectar secretion and subsequent low honey yield. Partitioning of resources between the vegetative and reproductive functions has been described as a distinct adaptation of plant species found in dry climatic conditions (Singh and Kushwaha, 2006).

According to the vegetation composition analysis, the dominant species in most sites include: Z. spina-christi, A. tortilis, A. ehrenbergiana, A. etbaica, Acacia origena, Acacia johnwoodii and Anisotes trisulcus contributing for 71.57% of the total studied bee forage coverage of the areas. The distribution of these species was either mixed or pure stands with relatively low species diversity with Shannon’s species diversity indices (H’) of 0.5–1.52 for different sites. The average vegetation cover values of the study areas were low (<30%). The relative density, frequency, species important values and relative apicultural values of some major species of the study area are shown in (Table 1). The highest relative frequencies of 0.48, 0.31 and 0.29 were recorded for A. etbaica, A. tortilis Z. spina-christi respectively indicating these species are commonly found in most of the studied plots. The same species have relatively better cover values (Table 1) indicating the species are more dominant in the region.

Causes for sparse vegetation coverage and limited plant diversity of the area could be due to many interrelated factors like: high moisture stress, human interference, recurrent drought, soil erosion, overgrazing, limited efforts toward conservation and rehabilitation. Extensive human interferences and climate changes and subsequent deterioration and continuous decline of many tree species and forest habitats have been reported for the region (Batanouny, 1991; Hall, 2005; El-Juhany, 2009; Hall et al., 2010).

According to the apicultural value estimation, the major plants in order of importance are Z. spina-christi, A. johnwoodii, A. ehrenbergiana, A. tortilis, A. origena and A. etbaica (Table 1). It was found that there was no positive significant correlations (r = -0.005, P = 0.493, N = 14) between IVI value and apicultural value (AV) of the species, which could be because IVI value is mainly related to the timber value while the AV value depends on the number of flowers per plant and nectar sugar amount per flower. Because of its very intermittent flowering nature; the honey from A. johnwoodii is not commonly exploited in the area. Moreover, despite its high nectar secretion potential, A. ehrenbergiana is also another bee forage plant which is not well exploited for honey production. This may be due to the co-existence and concurrent-flowering of the species with A. tortilis and the more

| Species                  | Ecology         | flowering and migration periods |
|--------------------------|-----------------|--------------------------------|
| Dipterygium glaucum     | coastal plain   | Dec.-Jan.                      |
| Ziziphus spinachristi   | Western midland | Oct.-Nov.                     |
| A. tortilis & A. ehrenbergiana | Western midland | Feb.-Mar.                    |
| Acacia asak             | Western slopes  | Aug.-Sep.                     |
| Lavendula spp           | Mountain tops   | May-June                      |
| Acacia origena          | Mountain tops   | May-June                      |
| Ziziphus spinachristi   | Eastern midland | sep.-Oct.                     |
| A. tortilis & A. ehrenbergiana | Eastern midland | Mar.-May                     |

Fig. 2. Schematic representation of the spatio-temporal distribution of major bee forages of the study area and seasonal colony migration.
preferences of honey bees toward *A. tortilis* flowers than *A. ehrenbergiana* (pers. obs. Nuru). The low preferences of the honey bees may depend on the morphology of the *A. ehrenbergiana* flowers which consists relatively very dense and longer florets which may reduce the honey bees to access its nectar. The presence of variations in the preferences of honey bees toward different plant species as a result of morphology of flowers and chemistry of nectar has been documented (Cook et al., 2003; Duffield et al., 2008).

The application of GIS system is found to be useful to spatially present the measured plant with their features as layer on the satellite image of the plots Fig. 3. Such spatial presentation of the measured plants with their features in their respective geographical position is useful to monitor the vegetation and land use changes of the studied valleys over the course of time. The importance of application of GIS and Remote Sensed Satellite Image analysis in monitoring vegetation cover changes has been well demonstrated by Lambin and Ehrlich (1997), Shalaby and Tateishi (2007), Abou-Shaara (2013). However it was noticed that slight shifting of the GPS co-ordinates recorded points (plant as object) from the corresponding points from the satellite image (Fig. 3C).

### 3.2. Satellite image processing

By applying the Hopfield neural network clustering algorithm techniques to the raw satellite images of the area (Fig. 4A); it has been achieved to segment and show the general vegetation cover and distribution of the study area (Fig. 4B, green). Using the pixel labels from the satellite images of the vegetation biomass of study area it was possible to further classify the image into two discrete clusters (black and red) (Fig. 4C) based on their mean color values. Finally, we were able to correlate the different cluster colors obtained as a result of segmentations of the satellite image with our GIS based ground inventory records of two co-existing species (*A. tortilis* (black), and *A. ehrenbergiana* (red)) (Fig. 4C), that dominantly grow in the target areas. This color mean value correlation can be used to extrapolate the spatial distribution of the different bee tree species of the target areas. The possibilities of fine tune mapping of vegetation cover to the species level based on appropriate satellite image processing techniques has been well indicated (Gould, 2000; Rocchini et al., 2007; Xie et al., 2008; Sammouda et al., 2013).

This will help estimate area coverage of the different species, the honey production potential and the expected flowering period of the different species and helps to guide migratory beekeepers for successful utilization of the available resources in different times and places. The significance of inventorying of bee forage plants; in suitable apiary site selection, stock level determination and creation of awareness in protecting and conserving of the bee plants species and their general ecosystem have been well documented (Debisa and Admasu, 2009; Taha, 2015).

![Table 1](image-url)

**Table 1** Species importance and relative apicultural values of some major honey bee plant species of the study area.

| Species name       | Relative canopy cover/species | Relative density | Relative freq. | Species importance val.(IVI) | Average nectar sugar/tree (kg) | Relative apiculture value |
|--------------------|-------------------------------|------------------|---------------|-------------------------------|--------------------------------|---------------------------|
| Acacia asak        | 0.02                          | 0.01             | 0.03          | 0.06                          | 0.14                           | 0.01                      |
| Acacia ehrenbergiana| 0.20                          | 0.21             | 0.27          | 0.68                          | 0.55                           | 0.38                      |
| Acacia ebaraica    | 0.54                          | 0.76             | 0.48          | 1.79                          | 0.13                           | 0.23                      |
| Acacia gerrardi    | 0.05                          | 0.07             | 0.11          | 1.55                          | 0.11                           | 0.17                      |
| Acacia oerfata     | 0.01                          | 0.02             | 0.03          | 0.06                          | 0.04                           | 0.00                      |
| Acacia origena     | 0.21                          | 0.17             | 0.19          | 0.54                          | 0.50                           | 0.27                      |
| Acacia johwoodii   | 0.11                          | 0.04             | 0.10          | 0.26                          | 2.14                           | 0.56                      |
| Acacia tortilis    | 0.45                          | 0.41             | 0.31          | 1.16                          | 0.28                           | 0.33                      |
| Anisotes trisulcus | 0.06                          | 0.25             | 0.18          | 0.49                          | 0.11                           | 0.05                      |
| Lavandula dentata  | 0.10                          | 0.49             | 0.15          | 0.74                          | 0.12                           | 0.09                      |
| Lavandula pubescens| 0.11                          | 0.44             | 0.17          | 0.81                          | 0.11                           | 0.09                      |
| Nepeta deflersiana | 0.13                          | 0.22             | 0.21          | 0.66                          | 0.12                           | 0.08                      |
| Otostegia fruticosa| 0.12                          | 0.44             | 0.12          | 0.71                          | 0.11                           | 0.08                      |
| Z. spina-christi    | 0.29                          | 0.10             | 0.29          | 0.68                          | 0.70                           | 4.52                      |

*Apicultural value of the species was calculated by multiplying the species important value index with average nectar sugar amount per tree.*

![Fig. 3](image-url)

**Fig. 3.** (A), Showing one of the studied valley with its plots as layer on the satellite image of the area; (B) one of the measured plot with trees as points and (C) focuses of the plot features overlapping with the satellite image of the plots.
4. Conclusion

Despite arid conditions of the area very diverse honey bee forages were recorded in the study areas. However the major honey source species are very few and are mainly trees and shrubs. The major bee forage species of the target area have been found to distribute spatiotemporally into eight distinct groups. These allow beekeepers to exploit the resources at different places and seasons. The application of ground inventory work supported with GIS system and analysis of Remote Sensed Satellite Image of vegetation of an area would be better approach to spatially and temporally map the bee forage resource of an area for their efficient utilization.

Acknowledgments

The project was financially supported by King Saud University, Saudi Arabia Vice Deanship of Research Chairs.

References

Abd-ElGhani, M.M., 1997. Phenology of ten common plant species in western Saudi Arabia. J. Arid Environ. 35, 673–683.
Abou-Shaara, H.F., 2013. Using geographical information system (GIS) and satellite remote sensing for understanding the impacts of land cover on apiculture over time. Int. J. Remote Sens. Appl. 3 (4), 171–174.
Abou-Shaara, H.F. 2015. Potential honey bee plants of Egypt. Cercetări Agronomice In Moldova, vol. XLVIII, No. 2 (162).
Abou-Shaara, H.F., 2015b. Pollen sources for honey bee colonies at land with desert nature during deathr period. Agronomical Res. Moldavia 48 (3), 73–80.
Admasu, A.M., 2003. Botanical Inventory and Phenology in Relation to Foraging Behaviour of the Cape Honeybees (Apis mellifera capensis) at a Site in the Eastern Cape (Master thesis). Rhodes University, South Africa.
Al-Ghamdi, A.A., 2007. Beekeeping and honey production in Saudi Arabia. Fifth Conference of Arab Beekeepers Association, the Federation of Arab beekeepers, Libya, Tripoli.
Amoko, J., 1997. Apiculture in Ghana the use of palynology to determine the Renewable resources exploited by honeybees Apismelliferadansonsii PhD thesis University of Wales. College of Cardiff, Cardiff, UK approach, West Sussex, England, pp. 96–97.
Baraouny, K.H., 1991. Vegetation of the Summan (Arabia); Pattern and process as affected by human impact and modern technology. In: Proceedings of IVth International Rangeland Congress, Montpellier, France, 4, pp. 310–314.
Borchert, R., 1994a. Soil and stem water storage determine phenology and distribution of tropical dry forest trees. Ecology 75, 1437–1449.
Borchert, R., 1994b. Soil and stem water storage determine phenology and distribution of tropical dry forest trees. Ecology 75, 1437–1449.
Chaudhary, S. (Ed.), 2001. Flora of the Kingdom of Saudi Arabia. Ministry of Agriculture & Water, Riyadh. Vols. 1–3.
Chemas, A., Rico-Gray, V., 1991. Apiculture and management of associated vegetation by the Maya of Tixcacaltuyub, Yucatán, México. Agrofor. Syst. 13, 13–25.
Collenette, I.S., 1999. Wildflowers of Saudi Arabia. National Commission for Wildlife Conservation, Riyadh. p. 799.
Cook, S.M., Awmack, C.S., Murray, D.A., Williams, I.H., 2003. Are honey bees’ foraging preferences affected by pollen amino acid composition? Ecol. Entomol. 28 (5), 622–627. http://dx.doi.org/10.1046/j.1365-2311.2003.00548.x.
Crane, E., 1990. Bees and Beekeeping: Science, Practice and World Resources. Oxford Heinemann Newnes.
Debisa, L., Admasu, A., 2009. Identification and Evaluation of Bee Flora Resources in Arid and Semiarid Agro-ecological Zones of South East of Oromia. Annual research directory report Holeta Bee Research Center, Ethiopia. p. 65.
Duffield, G.E., Gibson, R.C., Gilhooley, P.M., Hesse, A.J., Inkley, C.R., Gilbert, F.S., Barnard, C.J., 2008. Choice of flowers by foraging honey bees (Apis mellifera); possible morphological cues. Ecol. Entomol. 18 (3), 191–197. http://dx.doi.org/10.1111/j.1365-2311.1993.tb01083.x.
El-Juhany, L.I., 2009. Forest degradation and potential rehabilitation in southwest Saudi Arabia. Aust. J. Basic Appl. Sci. 3 (3), 2677–2696. <http://faculty.ksu.edu.sa/69123/pdf> accessed December, 2012.
Feinsinger, P., Beach, J.H., Linhart, Y.B., Busby, W.H., Murray, K.G., 1987. Disturbance, pollinator predictability, and pollination success among costa rican cloud forest plants. Ecology 68 (5), 1294–1305. http://dx.doi.org/10.2307/1939214.
Felzenszwalb, P.F., Huttenlocher, D.P., 2004. Efficient graph-based image segmentation. Int. J. Comput. Vision 59, 167–181.

Flombaum, P., Sala, O.E., 2007. A non-destructive and rapid method to estimate biomass and aboveground net primary production in arid environments. J. Arid Environ. 69 (2), 352–358.

Fox, G.A., 1990. Drought and the evolution of flowering time in desert annuals. Am. J. Bot. 77, 1508–1518.

Gould, W., 2000. Remote sensing of vegetation, plant species richness, and regional biodiversity hotspots. Ecol. Appl. 10 (6), 1861–1870. http://dx.doi.org/10.1890/1051-0746(2000)010[1861:RSVPS]2.0.CO;2.

Hampson, G.S., Al-Ghamdi, A.A., Ismaiel, S., Al-kahtani, S., Yilma, T., Ansari, S., Lambin, E.F., Ehrlich, D., 1997. Land-cover changes in sub-saharan Africa (1982–2000). Philos. Trans. R. Soc. Lond., B 352 (1334), 1175–1188.

Hall, M., Coker, P., 1992. Vegetation Description and Analysis. Belhaven Press, London, p. 363.

Kim, M.S., Kim, S.H., Han, J., Kang, M.S., Park, Y.K., 2011. Honeybee visit and nectar production of Cynanchum japonicum. J. Apic. Sci. 56, 249–259. http://dx.doi.org/10.1111/j.1442-9993.2010.01010.x.

Kent, M., Coker, P., 1992. Vegetation Description and Analysis. Belhaven Press, London, p. 363.

Khanbash, M.S., 2001. Conservation of zizyphus trees, from Deterioration to Raise Honey Productivity and Maintain its Quality. A Study introduced to Fund box to encourage agricultural production and fisheries of the Republic of Yemen, p. 59.

Kim, M.S., Kim, S.H., Han, J., Kang, M.S., Park, Y.K., 2011. Honeybee visit and nectar secretion characteristics of the Chinese Hawthorn Crataegus pinnatifida Bunge. J. Apic. Sci. 26 (1), 11–14.

Lambin, E.F., Ehrlich, D., 1997. Land-cover changes in sub-saharan Africa (1982–1991): application of a change index based on remotely sensed surface temperature and vegetation indices at a continental scale. Remote Sens. Environ. 61 (2), 181–200.

Mallick, S.A., 2000. Technique for washing nectar from the flowers of Tasmanian leatherwood (Eucryphia lucida Eucryphiaceae). Aust. Ecol. 25 (2), 210–212. http://dx.doi.org/10.1046/j.1442-9993.2000.001010.x.

Nur, A., Awad, A.M., Al-Ghamdi, A.A., Alqarni, A.S., Radloff, S.E., 2012. Nectar of Nuru, A., Awad, A.M., Al-Ghamdi, A.A., Alqarni, A.S., Radloff, S.E., 2006. Above and belowground ecosystem biomass and carbon pools in an age-sequence of temperate pine plantation forests. Agric. For. Meteorol. 140 (30), 51–63.

Peichl, M., Arain, M.A., 2006. Above and belowground ecosystem biomass and carbon pools in an age-sequence of temperate pine plantation forests. Agric. For. Meteorol. 140 (30), 51–63.

Pérez, R., 2004. Allometry and partitioning of above- and belowground tree biomass in an age-sequence of white pine forests. Ecol. Manage. 251 (1–3), 68–80.

Peñuelas, J., Filella, I., Comas, P., 2002. Changed plant and animal life cycles from 1952 to 2000 in the Mediterranean region. Glob. Change Biol. 8, 531–544.

Peñuelas, J., Filella, I., Zhang, X., Llorens, L., Ogaya, R., Lloret, F., Comas, P., Estiarte, M., Terradas, J., 2004. Complex spatiotemporal phenological shifts as a response to rainfall changes. New Phytol. 161, 837–846. http://dx.doi.org/10.1111/j.1469-8137.2003.01003.x.

Rathcke, B., 1988. Flowering phenologies in a shrub community: competition and constraints. J. Ecol. 76, 975–994.

Rocchini, D., Ricotta, C., Chiarucci, A., 2007. Using satellite imagery to assess plant species richness: the role of multispectral systems. Appl. Veg. Sci. 10, 325–332.

Sahoo, P.M., Rai, A., Singh, R., Handique, B.K., Rao, C.S., 2005. Integrated approach based on remote sensing and GIS for estimation of area under paddy crop in North-Eastern hilly region. J. Indian Soc. Agr. Stat. 59 (2), 151–160.

Samhouda, R., Nur, A., Ameur, T., Al-Chamdi, A., 2013. Agriculture satellite image segmentation using a modified artificial hopfield neural network. Comput. Hum. Behav. 30 (2014), 436–441. http://dx.doi.org/10.1016/j.chb.2013.06.025.

Shalaby, A., Tateishi, R., 2007. Remote sensing and GIS for mapping and monitoring land cover and land-use changes in the Northwestern coastal zone of Egypt. Appl. Geogr. 27 (1), 28–41.

Shi, J., Mallick, J., 2000. Normalized cuts and image segmentation. IEEE Trans. Pattern Anal. Mach. Intell. 22, 888–905.

Singh, K.P., Kushwaha, C.P., 2006. Diversity of flowering and fruiting phenology of trees in a tropical deciduous forest in India. Ann. Bot. 97 (2), 265–276.

Stone, G.N., Willmer, P.G., Rowe, J.A., 1998. Partitioning of pollinators during flowering in an African Acacia community. Ecology 79, 2808–2827.

Suganuma, H., Abe, Y., Taniguchi, M., Tanouchi, H., Utsugi, H., Kojima, T., Yamada, K., 2006. Stand biomass estimation method by canopy coverage for application to remote sensing in an arid area of Western Australia. For. Ecol. Manage. 222 (1–3), 75–87.

Taha, E.A., 2015. A study on nectar and pollen sources for honeybee, Apis mellifera L. in Al-Ahsa Saudi Arabia. J. Entomol. Zool. Stud. 3 (3), 272–277.

Williams-Linea, G., 1997. Phenology of deciduous and broad leaf ever-green tree species in a Mexican tropical flower montane forest. Global Ecol. Biogeogr. lett. 6, 115–127.

Wulder, M.A., White, J.C., Han, T., Coops, N.C., Cardille, J.A., Holland, T., Grills, D., 2008. Monitoring Canada’s forests. Part 2: national forest fragmentation and pattern. Can. J. Remote Sens. 34 (6), 563–584.

Xie, Y., Sha, Z., Yu, M., 2008. Remote sensing imagery in vegetation mapping: a review. J. Plant Ecol. 1 (1), 9–23. http://dx.doi.org/10.1093/jpe/itm005.