Implementing Cost-effective Co-treatment of Domestic and Food-industrial Wastewater by Novel Methods for Estimating Industrial Load

Vince Bakos*, Péter Szombathy1, József Simon1, Andrea Jobbágy1

1 Department of Applied Biotechnology and Food Science, Faculty of Chemical Technology and Biotechnology, Budapest University of Technology and Economics, H-1111 Budapest, Műegyetem rkp. 3., Hungary
* Corresponding author, e-mail: vbakos@mail.bme.hu

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Abstract
Detection and characterization of hidden industrial inflows causing high fluctuations of the inlet load, is a challenging issue pushing plant operators for a cost-effective solution at regional wastewater treatment plants (WWTPs). On the other hand, carbon source of food industrial origin may have a good use at WWTPs facing otherwise inlet carbon source deficiency.

In a case study of a regional domestic WWTP receiving seasonally organic carbon-rich discharge from a fruit juice factory, a new method combining on-site measurements and mathematical modelling was developed and successfully applied for estimating the quality and quantity of both industrial influent load and incoming domestic wastewater streams properly. The originally un-staged bioreactor system operated at low dissolved oxygen (low DO) concentration was unable to meet effluent nitrogen requirements with an additional constant risk of encouraging filament growth. A novel screening method based on special sampling campaigns for estimating carbon availability and C:N ratios of influent wastewater streams coming separately from the large catchment area, was developed and applied. Staging of the previously low DO basins into a flexible system containing non-aerated selectors proved to be efficient for enhancing both biological nitrogen removal and sculpturing appropriately settling biomass.

Keywords
activated sludge, carbon source availability, wastewater co-treatment, mathematical modelling, estimation of industrial influent

1 Introduction
Influents of Hungarian domestic wastewater treatment plants (WWTPs) are generally poor in readily biodegradable organic carbon source. However, depending on local water consumption practice, in retention time in sewer network as well as possible additional industrial discharge, remarkable differences can be measured from site to site [1, 2]. Although low dissolved oxygen (low DO) environment may be favorable for cost-effective removal of nitrogen [3], it may lead to performance deterioration [4, 5] and cause enhanced N₂O emission as well [6]. Additionally, low substrate (low S) concentration and/or low DO conditions may also increase the risk of excessive growth of filaments resulted in poor biomass settleability [7, 8]. In numerous cases, efficient biological nitrogen removal cannot be achieved without external organic carbon source dosing also by further use of liquid wastes for denitrification [9–12]. Carbon shortage of domestic influent wastewater has been becoming a global challenge in the last decade [2, 13–18]. Therefore, the cost-effective new technology of applying floating seals for excluding undesired oxygen penetration from non-aerated selectors [8, 19] is required as well as the further use of nutrients in agriculture is encouraged [20] for sustainable wastewater treatment and source recovery. At regional WWTPs with large catchment areas huge fluctuations may happen in influent quantity and quality including influent C:N ratios [2, 21], especially when industrial wastewater flows are also discharged into the municipal sewer network [22]. Food-industrial wastewater streams are generally readily biodegradable, however, their C:N ratios may be very different depending on the types of industry. Several kinds of food-industrial wastewaters (i.e. deriving from wineries, fruit juice, soft drink and beverage productions, etc.) may be severely nutrient deficient. In these cases, besides the remarkable operational costs, external nutrient dosing may have several
risks and uncertainties [23]. Therefore, advanced technologies could be preferred, such as the pioneering development and application for encouraging growth of GAOs (Glycogen Accumulating Organisms) as cost-effective solution [24–26]. In wastewater treatment plants, receiving both carbon deficient municipal and carbon-rich (but nutrient deficient) food-industrial discharges from heterogeneous catchment areas, co-treatment of these influents may be a powerful solution resulting in higher denitrification capacity [22, 27–29] as well as other innovative co-treatment solutions may also lead to cost-effective operation [30, 31].

Case study carried out at a Hungarian, regional municipal activated sludge WWTP operated at low DO conditions and receiving seasonally high carbon content wastewater from fruit juice industry (DOM&IND-WWTP, hydraulic capacity: 1100 m$^3$ d$^{-1}$) aimed to upgrade biological nitrogen removal and decrease the risk of filamentous bulking as well. The quality and quantity of diverse wastewater streams collected from the large and heterogeneous catchment areas were not appropriately known. Due to the high hydraulic retention time (HRT ≈ 1.5 d) in the activated sludge (AS) basins, industrial influent was noticed mainly by delayed detection of remarkable changes in plant performance. Therefore, the project for upgrading required the development of an appropriate methodology for detecting and quantifying the incoming industrial load as well as exploring and characterizing the diverse influent wastewater streams.

2 Investigated full-scale system and research concept

The original activated sludge wastewater treatment technology of the DOM&IND-WWTP consisted of four aeration basins connected in parallel as illustrated in Fig. 1. Dissolved oxygen (DO) concentration in the bioreactors was maintained below 1 mg L$^{-1}$, typically in the range of 0.1–0.5 mg DO L$^{-1}$ in order to ensure simultaneous nitrification and denitrification. The plant receives domestic wastewater transported through a short sewer system (from the city nearby), and an additional part of domestic wastewater streams coming from higher distances (15–20 km) of two agglomeration zones as well as seasonal industrial discharge of the fruit juice factory situated closer to the city. The mixed influent load had high fluctuations resulted in a broad influent quality and quantity spectrum.

The fully aerobic, but low DO system was unable to meet the effluent criteria (NH$_4$N 5 mg L$^{-1}$, TN 30 mg L$^{-1}$) at the relatively high Mixed Liquor Suspended Solids (MLSS) concentration of 5 g L$^{-1}$ applied, since depending on the inlet load either nitrification or denitrification failed. At the same time,

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**Fig. 1** Technological layout of the DOM&IND-WWTP with the indication of sampling points (see black numbers on grey labels) chosen for on-site concentration profile measurements.
periodically serious AS settleability problems occurred due to high filamentous abundance.

The upgrading concept aimed to transform the CSTR (Completely Stirred Tank Reactor) like, low-DO system into an appropriately staged, flexible technology through evaluation of operational data and careful investigation of incoming wastewater streams using on-site measurements and mathematical modelling.

3 Materials and methods

The operational and analytical data (temperature, Sludge Volume Index - SVI, Mixed Liquor Suspended Solids - MLSS, inlet and outlet parameters: pH, Chemical Oxygen Demand - COD, five-day Biochemical Oxygen Demand - BOD$_5$, Total Suspended Solids - TSS, Total Organic Carbon – TOC, Dissolved Organic Carbon – DOC, Total Kjeldahl Nitrogen – TKN, NH$_4$N, NO$_2$N, NO$_3$N and Total Phosphorus – TP concentrations) of the previous 3 years were carefully evaluated and discussed with the operator in order to explore former operational nuisances and their possible causes.

On-site concentration profile measurements were carried out both on the sewer network and at the WWTP (see Fig. 1) four times, during the period from March to August (i.e. March 9$^{th}$, March 21$^{st}$, June 14$^{th}$ and August 7$^{th}$). On each selected day of sampling four campaigns were carried out at all sampling points shifted in time consecutively (i.e. early morning, late morning, early afternoon and evening). Temperature, SVI, bioreactor DO concentration profile and MLSS as well as inlet and outlet pH, COD, dissolved COD, BOD$_5$, TSS, TOC, DOC, NH$_4$N, NO$_2$N, NO$_3$N, TKN, dissolved TKN, PO$_4$P and TP concentrations were measured. For determining dissolved parameters samples were centrifuged (12 000 rpm, 5 min) and filtered through filter membrane (pore size: 0.45 μm), then analysis was carried out from the filtrate. AS samples were investigated under microscope (Olympus CX4i) on each sampling day. Dissolved oxygen (DO) concentration and pH were measured by a portable WTW Multi 350i meter (WTW GmbH, Weilheim, Germany), for BOD$_5$ manometric method and WTW OxiTop Control System was applied (WTW GmbH, Weilheim, Germany). Generally, the parameters were measured according to the Hungarian Standards which matches international protocols [32]. For bio-kinetic calculations ASM1 (Activated Sludge Model 1) based SSSP [33] and ASM2d based BioWin 4.1 [34] softwares were used.

4 Results and discussion

As it is illustrated in Fig. 2 evaluation of plant data showed that nitrification efficiency dropped significantly according to the Hungarian Standards which matches international protocols [32]. For bio-kinetic calculations ASM1 (Activated Sludge Model 1) based SSSP [33] and ASM2d based BioWin 4.1 [34] softwares were used.

| Parameters/ Wastewater origin | T [°C] | COD [mg L$^{-1}$] | Diss. COD [%] | Diss. COD / Tot. COD ratio [%] | Diss. TN [mg L$^{-1}$] | Diss. COD / TN ratio [-] |
|-------------------------------|--------|------------------|--------------|------------------------------|----------------------|------------------------|
| Short sewer (city and fruit juice factory) | 16.9 ± 1.5 | 795 ± 70 | 558 ± 29 | 70 ± 4.8 | 48.0 ± 8.8 | 11.6 ± 2.1 |
| Long sewer (urban agglomeration) | 10.1 ± 0.4 | 338 ± 55 | 118 ± 23 | 35 ± 3.9 | 91.1 ± 13.1 | 1.3 ± 0.2 |
down (see Fig. 2 (a)) in accordance with increased influent organic load (see Fig 2 (b)) even in the indicated range when bioreactor temperature was increasing. It can be assumed that during this period dissolved oxygen (DO) concentration decreased due to the increasing organic load and decreasing solubility of oxygen.

At the same time, increased denitrification efficiency can be observed (see Fig. 2 (a)) in accordance with the assumed low DO conditions and the high availability of carbon source (see Fig. 2 (b)). Nevertheless, as it is shown in Fig. 2 (c) SVI level could not be reliably stabilized below 150 cm$^3$g$^{-1}$ except for chemical dosing the data refer to. On the other hand, in the cold period (from January through early May) there could be full nitrification and carbon removal whereas decreased denitrification efficiency detected, most probably, due to an increased DO level.

Through evaluation of plant data the causes of treatment nuisances could not be clearly explored and neither the responsible process failures nor the influencing influent wastewater streams could be satisfactorily identified.

On-site measurements showed huge differences regarding the quality of the diverse incoming wastewater streams (see Table 1). Although the mixed wastewater of the city and the fruit juice factory proved to be appropriate for efficient denitrification, the wastewater of the distant agglomeration zones proved to be severely carbon deficient.

Although the on-site measurements provided a detailed screening of the diverse incoming wastewater streams, additional background information received from the operator about the industrial discharge around the sampling days was rather uncertain, and inflows coming by night or suddenly as shock-load could not be detected either. Regarding these uncertainties and taking also into account the high HRT (~1.5 d) in the AS basins, on-site profile measurements did not prove to be satisfactory for fitting the measured effluent concentrations to an appropriate influent quality. The only reliable information was that there was no industrial discharge before the sampling on August 7$^{th}$ (last sampling day of the campaign) due to the production stoppage of the fruit juice factory. However, the rate of industrial discharge was unknown for the other three dates of the campaign.

For developing a good estimation, the typical characteristics of different domestic wastewater streams were determined based on the concentration profile measurement data detected on August 7$^{th}$. Then, further calculations were carried out applying mixing equations in order to estimate the industrial wastewater quality with the use of data acquired during the previous sampling days. Finally, a wastewater load spectrum (both for quality and relatable quantity) was built up for finding the one entering the plant most likely (see Table 2 and Fig. 3).

In order to identify the industrial fraction in the mixed influents of the first three sampling days (March 9$^{th}$, March 21$^{st}$, and June 14$^{th}$), a new method was developed and applied for influent wastewater characterization including both on-site concentration profile measurements and application of ASM modelling. Mathematical simulation characterizing the low DO system was carried out for the whole wastewater loading spectrum determined previously on the basis of analytical and influent flow data (see Table 2 and Fig. 3) as well as the technological and design data presented in Fig. 1. Fig. 4 illustrates the curves of the calculated effluent NO$_3$N concentrations in case of different influent wastewater qualities investigated. The detected effluent nitrate concentrations are indicated by large rectangular filled black markers on the diagram of Fig. 4. Data of August 7$^{th}$ were used as

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**Table 2** Characterization of potentially occurring mixed influent wastewater streams entering the plant regarding the flow rates presented in Fig. 3.

| Parameter | Tot. COD [mg L$^{-1}$] | Diss. COD [mg L$^{-1}$] | Diss. COD / Tot COD [%] | TN [mg L$^{-1}$] | Diss. TN [mg L$^{-1}$] | NH$_4^+$N [mg L$^{-1}$] | Diss. COD / Diss. TN [-] |
|-----------|----------------------|------------------------|-------------------------|----------------|-----------------|----------------|----------------|
| Purely domestic without industrial fraction | 551 | 260 | 47 | 90 | 80 | 70 | 3.3 | Poor |
| Domestic + low industrial discharge | 621 | 345 | 56 | 87 | 78 | 68 | 4.4 | Marginal |
| Domestic + moderate industrial discharge | 737 | 489 | 66 | 83 | 75 | 64 | 6.5 | Good |
| Domestic + full industrial discharge | 802 | 569 | 71 | 80 | 73 | 62 | 7.8 | Good |
| Domestic + increased industrial discharge | 938 | 735 | 78 | 75 | 69 | 57 | 10.7 | Excellent |
a reference for model fitting as there was no industrial influent in that period due to production break in the fruit juice factory. (The same calculations were carried out for effluent NH$_4$N concentrations as well.) Good estimations for targeting influent industrial wastewater fraction of the first three sampling days could be carried out through the comparison of the calculated nitrate concentrations (curves) and the measured effluent values (large rectangular filled black markers).

The mathematical simulation study confirmed the hypothesis that the original, low DO system is not able to fulfill the effluent requirements, neither in case of lacking or low industrial influent (leading to high effluent nitrate) nor in case of high industrial discharge (due to deterioration of nitrification at lower temperatures).

Therefore, a flexible, staged bioreactor arrangement containing separated anoxic and aerated basins was proposed instead of the previously operated fully aerated, CSTR-like low DO system as illustrated in Fig. 5. Aerated low DO basins were reconfigured into tanks in series, and one of them was staged for an anoxic and an alternatively operated anoxic and aerobic basin at the head of the AS system. While in case of lacking or low industrial influent anoxic mode was recommended for the alternative bioreactor, at higher industrial discharge aeration was proposed. In the aerated bioreactors DO concentration was suggested to set high enough (up to ~2 mg L$^{-1}$) for efficient nitrification in the reduced, aerated volume. Possibility of internal nitrate recirculation was added in order to enhance pre-denitrification depending on inlet carbon source availability.

The calculated effluent ammonium concentrations of the staged system showed full nitrification in the investigated temperature and influent load ranges as illustrated in Fig. 6. The reduced aerated volume (with decreased aerobic SRT – Sludge Retention Time) was, however, operated at the higher DO level and thereby proved to be enough for efficient ammonium oxidation. The effluent ammonium concentration increased above 1 mg NH$_4$N L$^{-1}$ only in case of increased industrial discharge at 12 °C but remained under 3 mg NH$_4$N L$^{-1}$ without any risk of exceeding the limit value of 5 mg NH$_4$N L$^{-1}$.

In case of industrial inflow effluent nitrate concentration could be safely maintained under 20 mg NO$_3$N L$^{-1}$ (see Fig. 7), however, without industrial discharge effluent concentrations were typically in the range of 22–27 mg NO$_3$N L$^{-1}$ referring to the poor readily biodegradable carbon source availability of the mixed, purely domestic wastewater. These results verified that the industrial load coming as the discharge of the fruit juice factory has beneficial effect on the influent C:N ratio, leading to enhanced denitrification efficiency.

However, in the absence of industrial discharge, the effluent TN may go up close to the limit value of 30 mg TN L$^{-1}$ (see Fig. 8) and external carbon dosing may be necessary for matching safely the effluent limit criteria. Results are in good accordance with earlier studies on co-treatment of domestic and dairy wastewater [21].

The new, staged and flexible configuration developed by onsite measurements and mathematical simulation for the originally CSTR-like system enabled to meet the effluent nitrogen criteria for the whole investigated inlet wastewater load spectrum in the bioreactor temperature range of 12–26 °C without external carbon source dosing and bioreactor volume extension. Calculations were carried out for estimating the achievable operation cost savings by the use of food-industrial wastewater for cost-effective co-treatment instead of dosing commercial external carbon source for denitrification. In case of treating purely

![Fig. 3 Potentially occurring flow portions of the mixed influent hydraulic load (different dissolved COD / dissolved TN ratios are indicated on the top of the columns of the diagram, changing between 3.3 and 10.7).](image-url)
Fig. 4 Measured (large rectangular filled black markers) and calculated (curves and related small data markers) effluent nitrate concentrations at different inlet loads with different amounts of industrial discharge.

Fig. 5 Technological layout of the upgraded (staged and flexible) system.
readily biodegradable carbon deficient domestic wastewater the necessary external carbon source dosing would be approx. 93 kg methanol d\(^{-1}\) at an average incoming domestic wastewater flow rate of 700 m\(^3\) d\(^{-1}\) for achieving an influent C/N ratio appropriate for efficient denitrification (as in case of Domestic + moderate industrial discharge wastewater quality in Table 2). At the current price of approx. 0.54 EUR kg\(^{-1}\) methanol available for industrial use in Hungary, the estimated additional yearly operation costs may achieve 18 500 EUR y\(^{-1}\) even at this small regional WWTP. On the basis of the onsite study a basically similar, flexible and staged system was implemented in the full-scale system that could efficiently control the growth of filaments as well.

5 Conclusions and major findings
For state of the art design and upgrading of WWTPs the plant and its catchment area should be considered as an integrated system, particularly in case of combined treatment of domestic and industrial discharges.

In the case study of a domestic activated sludge wastewater treatment plant receiving high carbon content wastewater of a fruit juice factory seasonally, a novel method has been developed for revealing inlet wastewater quality through combining on-site measurements with mathematical simulation. This new methodology can be applied for detecting occasional and/or seasonal influent loads of industrial origin, estimating the quantity of incoming industrial discharge, and discovering latent industrial (shock-)loads as well.

In case of highly fluctuating inlet quality, staged bioreactor arrangement and possibility of flexible operation proved to be much more efficient than the original un-staged, low S – low DO system aiming simultaneous nitrification and denitrification. Besides meeting the effluent nitrogen

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**Fig. 6** Simulation results for effluent NH\(_4\)N concentrations in case of staged bioreactor arrangement, including anoxic and aerobic basins.

**Fig. 7** Simulation results for effluent NO\(_3\)N concentrations in case of staged bioreactor arrangement, including anoxic and aerobic basins.

**Fig. 8** Simulation results for effluent TN concentrations in case of staged bioreactor arrangement, including anoxic and aerobic basins.
criteria the new bioreactor arrangement proved to be able to minimize the risk of filamentous bulking.

Application of appropriately staged bioreactor arrangement resulted in efficient carbon consumption coupled with enhanced biological nitrogen removal and facilitated good biomass settleability. The novel detection and quantification method for estimating influent industrial load and the proposed co-treatment technology in staged bioreactors proved to be cost-effective for replacing external carbon source dosing and saving remarkable additional operational expenses.

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