Emerging Parasitic Protists: The Case of Perkinsea

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The last century has witnessed an increasing rate of new disease emergence across the world leading to permanent loss of biodiversity. Perkinsea is a microeukaryotic parasitic phylum composed of four main lineages of parasitic protists with broad host ranges. Some of them represent major ecological and economical threats because of their geographically invasive ability and pathogenicity (leading to mortality events). In marine environments, three lineages are currently described, the Parviluciferaceae, the Perkinsidae, and the Xcellidae, infecting, respectively, dinoflagellates, mollusks, and fish. In contrast, only one lineage is officially described in freshwater environments: the severe Perkinsea infectious agent infecting frog tadpoles. The advent of high-throughput sequencing methods, mainly based on 18S rRNA assays, showed that Perkinsea is far more diverse than the previously four described lineages especially in freshwater environments. Indeed, some lineages could be parasites of green microalgae, but a formal nature of the interaction needs to be explored. Hence, to date, most of the newly described aquatic clusters are only defined by their environmental sequences and are still not (yet) associated with any host. The unveiling of this microbial black box presents a multitude of research challenges to understand their ecological roles and ultimately to prevent their most negative impacts. This review summarizes the biological and ecological traits of Perkinsea—their diversity, life cycle, host preferences, pathogenicity, and highlights their diversity and ubiquity in association with a wide range of hosts.

Keywords: Perkinsus, Parvilucifera, X-cell parasite, broad host range parasite, emerging diseases, severe Perkinsea infection, opportunistic parasite

INTRODUCTION

Parasitism is a key component in all ecosystems, playing a fundamental role at the population level and wider ecological scales. Although parasites play a key role in food web interactions (Bjorbækmo et al., 2020), their diversity, dynamics, and influence on ecosystems remain neglected (Lafferty et al., 2006). In marine ecosystems, global environmental sequencing studies, based on the analysis of the small subunit rRNA-encoding gene (SSU rDNA) have revolutionized our conception of the microbial food webs with the (re-)discovery of an undescribed diversity shaping a more complex and cryptic global interactive network (Lima-Mendez et al., 2015; Guidi et al., 2016). However, the vast majority of these putative parasitic organisms are still only identified by taxonomic
marker gene sequences and fundamental questions about host preference, range, and interactions remain unanswered.

Parasitic organisms are defined on the basis of their trophic characteristics and, therefore, by their host range, which allow or not the spread to a novel host species or into new biogeographical areas (Cleaveland et al., 2001; Poulin and Mouillot, 2003). The host range, defined by Lymbery (1989), considers the number of host species by a given parasite species. Two main categories can be distinguished by the extent diversity of hosts infected: the narrow host range (NHR) and the broad host range (BHR). Despite NHR parasites seem to slightly dominate the host–symbiotic networks as described by Bjørnækmo et al. (2020), this dichotomy between BHR and NHR parasites is not static, and events like host shift can lead NHR to become BHR parasites and vice versa (Agosta et al., 2010). The distinction between NHR and BHR does not seem to be clear-cut for the majority of the parasitic taxa. Therefore, Desverges et al. (2002) developed the non-specific index (NSI), which classifies parasites based on their host range and the relatedness between host organisms. This index was adapted by Šimková et al. (2006) as the specificity index (SI) based on cyprinid fish and Dactylogyrus spp. (monogenean gill flukes) model with the following categories: (1) strict specialist living on a single host species, (2) intermediate specialist living on two or more host species from the same genus, (3) intermediate generalist living on two or more non-congeneric species of the same terminal clade, (4) generalist living on different hosts belonging to one host family and (5) “true” generalist living on different host species of different families (Figure 1A). However, estimating host range is particularly challenging in the case of uncultivable parasitic microorganisms. Recent studies highlight the strong potential of molecular methods, such as PCR, to detect putative cryptic host–parasite associations [see review by Bass et al. (2015)]. However, these molecular methodologies only reflect the presence/absence of the genetic signature of the parasite within the host tissue and do not identify paratenic, reservoir, dead-end, or accidental hosts. Hence, the combination of molecular and microscopic techniques should provide a powerful tool to access the putative range of host–parasite associations. In this review, to avoid confusion, we will look on how the putative association between hosts and parasites has been estimated by focusing, if possible, on the combination of molecular and microscopic detection methodologies. Finally, we consider here the host–parasite association in terms of species richness following the SI, but also in terms of accumulated scientific knowledge on these host–parasite interactions and choose to define BHR parasite as an organism able to ensure its own survival through the colonization of divergent host species in the same ecosystem or following a translocation event.

In recent years, host–parasite interactions have been influenced by anthropogenic consequences (e.g., global warming, habitat degradation, mass extinction, and introduction of exotic host or parasitic species). This phenomenon is illustrated by parasitic species dissemination worldwide and the severity of infectious disease in time and space, which can lead to mortality events (Harvell, 1999; Daszak, 2000). Indeed, interactions between a host and its pathogens are not fixed in time and space. They depend both on the efficiency of the defense of the host (immune system or strategies) and on mechanisms used by parasites to bypass these defense systems to establish infection (Råberg et al., 2014). Indeed, environmental stressors (e.g., temperature, pollution) modify the immune system of the host, which increase their vulnerability to parasitism (Goedken et al., 2005; Morley, 2010). For example, in a metazoan host, P. marinus-infected oysters exposed to the antifouling agent tributyltin (TBT) showed an increase in prevalence and mortality (Anderson et al., 1996), or in a protist host, herbicide exposure could lead to increase in chytrid infections in phytoplankton populations (e.g., Van den Wyngaert et al., 2013).

In addition, the introduction of exotic species, described as a global phenomenon, is also considered as a main factor of alteration of local networks (Cohen, 1998; Galil, 2000; Goedken et al., 2005). Over the last two centuries, 37% of the first recorded species introduction phenomenon occurred in the last 50 years (1970–2014), and this trend is not slowing down (Seebens et al., 2017). In aquatic ecosystems, the unintentional introduction of exotic species has increased substantially with the globalization of economies: Haplosporidium nelsoni (hypothetical Livestock vector: Crassostrea gigas) translocated from Asia to the United States (USA) (Burreson et al., 2000) or Bonamia ostreae (hypothetical livestock vector: Ostrea edulis) translocated from the United States to Europe (Elston et al., 1986). Each day, hundreds of species are passively transferred across the oceans by sea transport (e.g., ballast water), pet trade, or stock exchange of organisms (Ruiz et al., 2000; Patoka et al., 2018). If the majority of pest introductions fail due to the hostility of the new ecosystem (Williamson and Fitter, 1996; de Montaudouin et al., 2001), some find a new niche for their development and reproduction (e.g., the absence of enemies, such as predators, parasites, and competitors). Moreover, parasitic microorganisms with a high potential to invade and quickly adapt to a new environment share several common traits as being r-strategist, a wide host range with high phenotypic plasticity, high dispersal capacities and genetic diversity (Sakai et al., 2001; Litchman, 2010). All these conditions form the “ecological roulette” described by Carlton and Geller (1993).

Hence, the introduction of exotic parasites could lead to the emergence of new pathologies with, in some cases, disastrous consequences for the local host populations. One of the most famous examples of an emerging disease is the “Dermo” disease caused by the virulent and invasive protist Perkinsus marinus (Perkinsea, Alveolata). This parasitic protist has been identified as responsible for the mortality events of oysters (Crassostrea virginica) in, for example, the Gulf of Mexico and the Chesapeake Bay (United States) (Burreson et al., 1994; Andrews, 1996). This microeukaryote belongs to the Perkinsea (syn. Perkinsids, Perkinsozoa) lineage Alveolata, which has long been ignored in aquatic ecosystems with the exception of the notifiable agents, P. olseni and P. marinus. Nowadays, four parasitic lineages of Perkinsina have been described from a wide variety of aquatic ecosystems: Perkinsidae, Parviluciferaceae, Xcellidae, and the severe Perkinsina infection (SPI) agent.

Due to a general lack of knowledge of the diversity and biology of these parasitic protists, our understanding of their impact on the structure and functioning of aquatic ecosystems remains
limited. Our objective here is to review the main discovery of described Perkinsea lineage, of which some representative organisms could be classified as putative BHR species with a strong capacity to become successful invasive species in a context of global change.

**Perkinsus spp., an Emerging Parasite of Mollusks**

In 1946, following a mass mortality event (MME) of oyster stocks in Louisiana (Gulf of Mexico, United States), Mackin et al. (1950) identified a protist present in host tissue sample as the causative agent. First affiliated to a fungal lineage and named *Dermocystidium marinum* (Mackin et al., 1950), electronic microscopy observations then showed morphological characteristics, such as the presence of a subpellicular membrane, micropores, and a conoid-like structure (zoospore stage) suggesting a re-affiliation of this protist into the Apicomplexa phylum (Alveolata) (Perkins and Menzel, 1967; Perkins, 1976). While doubts remained about the accuracy of this affiliation (Siddall et al., 1997), molecular phylogenies based on ribosomal and actin sequences revealed that *Perkinsus* species were more closely related to dinoflagellates than apicomplexans (Reece et al., 1997). In 2003, multiple protein phylogeny indicated that *Perkinsus* is an early branch in dinoflagellate lineage (see schematic representation in Figure 2A; Saldarriaga et al., 2003).

Nowadays, seven species are described within the genus *Perkinsus*: *P. marinus*, *P. olseni*, *P. gugwadi*, *P. chesapeakei*, *P. mediterraneus*, *P. honshuensis*, and *P. beihaiensis* (see Figures 1B, 2B for the phylogenetic classification of these parasites). However, only *P. marinus* and *P. olseni*, the etiological agents of “Dermo” disease and Perkinsosis, respectively, have significant negative impacts on mollusk populations worldwide (Siddall et al., 1997). These two infectious agents are today within...
FIGURE 2 | Major phylogenetic relationships within the Perkinsea and other alveolates. (A) Schematic radiation of Alveolata superphylum (not to scale) based on their rDNA phylogeny adapted from Chambouvet et al. (2020). The basal branch (dotted line) is hypothetical. (B) Maximum likelihood tree investigating the Perkinsea diversity based on 18S rRNA. The phylogeny was calculated from 134 taxa and 1,431 character alignment position. Seven sequences of dinoflagellates, Syndiniales, and MALV (marine alveolate) were used as an outgroup. ML bootstrap values (1,000 replicates, GTR+F+R5) and Bayesian posterior probability (8,000,000 generations, GTR+G+R5 model) were notated using the following convention: support values are summarized by black circles when ≥80%/0.9 and white circles when it is not the case but values ≥60%/0.6. When the topology is inconsistent in one of the inference methods, it is denoted by a ‘−’ (see phylogenetic analysis details in Supplementary SMM). The sequence origins are represented by squares of colors: light blue for marine waters, yellow for land waters, green for brackish waters, and red for wetland soil. Purple asterisks indicate host-associated sequences. The tree was annotated using Interactive Tree of Life (IToL) (https://itol.embl.de/, Letunic and Bork, 2019) and Inkscape (https://inkscape.org/en/).
the list of notifiable diseases in the World Organization for Animal Health (OIE) (Oie-Listed Diseases, 2021: OIE—World Organisation for Animal Health).

Some Perkinsus species have a putative BHR among mollusk species as, for example P. olseni infecting various clams (e.g., R. decussatus, R. philippinarum, Austrovenus stutchburyi, Tridacna maxima, T. crocea, and Pitar rostrata), oysters (Crassostrea rhizophorae, Saccostrea sp.), pearl oyster (e.g., Pinctada imbricata and P. fucata), and abalone (e.g., Haliotis rubra and H. laevigata) (Lester and Davis, 1981; Cremonete et al., 2005; Dungan et al., 2007; Sheppard and Phillips, 2008; Sanil et al., 2010; Ramilo et al., 2015; Pagenkopp Lohan et al., 2018), or P. chesapeakei, which can infect clams (R. decussatus, R. philippinarum) and oysters (C. rhizophorae, C. virginica) (Coss et al., 2001; Arzul et al., 2012; Dantas Neto et al., 2016).

Conversely other Perkinsus species show an intermediate host range as, for example, P. marinus infecting mostly oysters (e.g., C. agar, C. virgincia, C. agar, and Saccostrea palmula) (e.g., Enríquez-Espinoza et al., 2010; Cáceres-Martínez et al., 2012; da Silva et al., 2013, 2014).

Until now, most of the host species described above are of commercial interest but other bivalve and gastropod species may also be susceptible to infection, especially in the described geographical range. However, the Perkinsus putative wide host range established solely using microscopic and molecular methodology may also be overestimated. Indeed, these results do not allow the identification of dead-end hosts that prevent the parasite transmission to the definite hosts. Furthermore, despite their BHR character, host susceptibility needs to be considered to prevent and protect most sensible bivalve stocks. For example, the two genetically related oysters C. gigas and C. virginica respond with different susceptibility when challenged by P. marinus. Although infection is established in both hosts, only infection in C. virginica is lethal (Barber and Mann, 1994). Among Perkinsus species, different levels of prevalence depending on hosts species are recorded across the world (Figure 3). In the Chesapeake Bay, P. marinus and P. chesapeakei infect different species of oysters and clams living in sympathy. Among these hosts, P. marinus rarely infects clams compared with P. chesapeakei and exhibits oyster preference (Reece et al., 2008). Variability in pathogenicity across the BHR of Perkinsus spp. raises the question about the existence of specific strains. For example, such specific strain was recently highlighted for Perkinsinus marinus with the emergence between 1983 and 1990 of a new hypervirulent phenotypic strain that includes a shortened life cycle and a tropism shift from deeper connective tissues to digestive epithelia (Carnegie et al., 2021). Authors hypothesized that the development of this new strain may be related to reduced oyster abundance and the rapid establishment of the exotic parasite H. nelsonii in 1959. Hence, given the increase in epizootic diseases (= disease event in a non-human animal population analogous to an epidemic in humans) worldwide, consideration of the host range is absolutely essential because i) they may be the starting point for a host shift, ii) they represent vector of transmission with no apparent signs of disease, and iii) new native parasitic strains can emerge in response to biotic or abiotic stressors.

Perkinsus infection appears to occur directly by filtration without an intermediate host, with gills and labial palps playing a crucial role as entrance portal for the parasite (Chintala et al., 2002; Allam et al., 2013; Wang et al., 2018). Four different life stages (Figure 4A) are described today occurring inside or outside the host with an exception for P. aquwadi where all stages can be observed in host tissues (Blackburn et al., 1998). All of them can induce new infections into a healthy host. The trophozoite life stage proliferates by vegetative multiplication (palintomy) in the host tissue. Disruption of the cell wall allows the release of immature spherical trophozoites that will gradually enlarge becoming mature vegetative cells (Perkins, 1996). In case of heavy infection, trophozoites can invade totally host tissues, with occasional production of cutaneous white nodules (inflammatory reaction), and induce a global decrease in host fitness (e.g., lower filtration activity, retard in growth and reproduction) (Dittman et al., 2001; Lee et al., 2001; Choi and Park, 2010). When the host becomes moribund, the parasite enlarges and develops a thick cell wall becoming a hypnospor life stage (Valulis and Mackin, 1969; Perkins, 1996). When transferred to fresh seawater under favorable conditions of temperature and salinity (Auzoux-Bordenave et al., 1996), hypnosporos undergo multiple divisions leading to the formation of free-living biflagellate life stages, the zoospores, into a cellular structure called zoosporangium. At maturity, the zoosporangium releases zoospores in the medium allowing the infection of new hosts.

Perkinsus species are one of the most famous examples of successful invasive pests in marine environments (Figure 3A). Three different scenarios are described through three different Perkinsus species: (1) Expansion: P. marinus has shown a significant increase in its geographic area in the United States along the East coast by more than 500 km in 2 years (1990–1992). This expansion, which is now permanent, has been correlated with an increase in winter surface temperature due to milder winters (Ford, 1996; Ford and Smolowitz, 2007). (2) Introduction: P. chesapeakei appears to have been accidentally introduced in Europe via its vectors Mya arenaria, or the hard clam, Mercenaria mercenaria from the United States (Arzul et al., 2012). The sporadic detection of P. chesapeakei does not reflect an expansion dynamic. To date, no mortality event affiliated to P. chesapeakei has been recorded. (3) Introduction and expansion: P. olseni has been co-introduced in Europe with its host the Manila clam, Ruditapes philippinarum, in 1972 for the development of clam aquaculture (Ruano et al., 2015). After its first detection in the late 1980s, the first mortalities of clam stocks were attributed to this parasite at Ria de Faro in Portugal (Ruano and Cachola, 1986). Since then, P. olseni has produced mortalities in Spain, Portugal, and Italy, infecting both the exotic Manila clam, Ruditapes philippinarum, and the native European clam, Ruditapes decussatus (Azvedo, 1989; Figueras et al., 1992; Pretto et al., 2014). Following these first mortalities, P. olseni was detected along the French Atlantic coast with infection prevalence up to 100% (Lassalle et al., 2007), in contrast to the non-detection of Perkinsus spp. in most of (these) previously studied sites (Goggin, 1992). Similarly, histological studies conducted by Vilela (1951) on Minchinia...
FIGURE 3 | Geographical distribution of Perkinsea parasitic protist based on scientific literature from 1950 to 2020. Detection provenance of (A) Perkinsidae, (B) Parviluciferaeae, (C) Xcellidae family, and (D) SPI agent (detail of selected references in Supplementary Table 1) is indicated by a colored triangle when the parasite is detected simultaneously by molecular (qPCR or PCR) and microscopic methodologies (histology or RFTM incubation or cultures), and by a star when presence of the parasitic protists was linked to a mortality event. The color of triangles or stars designates the parasite species (see key). World map drawing is a free public domain vector cliparts (available at https://commons.wikimedia.org).
through their status as habitat engineers, as filter feeders, and as a connecting compartment between primary producers and predators (Jones et al., 1996; Coen and Grizzle, 2007; Gallardi, 2014). By cascade effect, parasitic infection can have deleterious effects on the different functions attributed to these populations: (i) by altering their bioturbatory activity and, thus, by modifying biogeochemical cycles at the water–sediment interface (Dairain et al., 2019), (ii) by impacting their filtration activity participating in the nitrogen cycle (= production of NH$_4^+$ leading to an increase in primary production) and helping to reduce turbidity (due to the increase in light penetration) (Gallardi, 2014), and (iii) by altering their physiological traits (e.g., growth, reproduction, and survival), which will modify the structure of habitat, biotic interaction, and the associated species richness (e.g., Thomas and Poulin, 1998). In addition to this infectious threat, benthic biodiversity is threatened by human activities (McCauley et al., 2015). Indeed, anthropogenic eutrophication and temperature increase are all factors that can weaken the immune system of the benthic organisms and lead to a decrease in their resistance to parasitic protists as Perkinsus spp. (Morley, 2010). Furthermore, the host range of these parasitic organisms might be wider than described in the scientific literature with the existence of other “non-commercial” potential hosts belonging to suborders infected by Perkinsus spp. Thus, even though many arguments allow classifying the Perkinsus species as a BHR parasite, we cannot determine the real consequences on stocks of “non-optimal” hosts especially when it is related to sporadic infections with low prevalence. Due to research focus on economically valuable species of mollusks like C. virginica or R. philippinarum, consequences of these infections on the global host range of Perkinsus spp. are still mainly undiscovered.

Parviluciferaceae, a Group of Parasites of Microalgae
Norén et al. (1999) identified small round structures abnormally present within the toxic dinoflagellate Dinophysis spp. on the west coast of Sweden. Molecular and microscopic analyses revealed that this parasitic organism was affiliated to the Perkinsids group and erected a new phylum, the Perkinsid Phylum (Norén et al., 1999). So far, five genera have been described and cultivated (Jeon et al., 2018). The first genus described was Parvilucifera genus encompassing P. infectans (Norén et al., 1999), P. multicavata (Jeon et al., 2018), P. sinerae (Figueroa et al., 2008), P. rostrata (Lepelletier et al., 2014b), P. corolla (Reñé et al., 2017b), P. catillosa, and Parvilucifera sp. (Alacid et al., 2020). Recently, three new genera were described, the genus Snorkelia containing P. prorocentri (Leander and Hoppenrath, 2008) renamed Snorkelia prorocentri (Reñé et al., 2017a), the genus Dinovorax including D. pyriformis (Reñé et al., 2017a), Tuberlatum encompassing T. coasi (Jeon et al., 2018), and finally Maranthos including M. nigrum (Reñé et al., 2021a). All these genera seem to have a basal position within the Parviluciferaceae group, with the exception of the Maranthos species, whose phylogenetic position is not yet clear (Jeon et al., 2018; Reñé et al., 2021a; Figures 1B, 2).

Members of the Parviluciferaceae family share many common traits including development and characteristic of their life cycle,
and their infection strategy (Reñé et al., 2017a). The infection begins with the entry of the zoospore into the dinoflagellate host cell (Figure 4B). Once inside, the parasite develops in the cytoplasm forming a spherical cell, the trophont (= trophocyte), which consumes its host for its own growth. At the end of the feeding stage, a free-living non-motile cell, the sporocyte (= immature sporangium), occupying the whole intracellular space, is released by breaking the theca of the host. Next, the cell undergoes morphological transformations into a spherical multinucleated dark structure called the sporangium (= mature sporocyte). The sporangium remains in dormancy until the detection of a host chemical signal (Garcés et al., 2013b). After a short maturation (~48 h), a sporangium with open aperture(s) releases several hundred new infectious cells (zoospores) (Alacid et al., 2017). To date, all described species are parasites of dinoflagellates including toxic species and possess a BHR (here belonging to a “true” generalist class) (Figueroa et al., 2008; Garcés et al., 2013a; Leon et al., 2018; Rodriguez and Figueroa, 2020; Reñé et al., 2021b) except S. prorocentri, isolated from Boundary Bay (Canada), which was known to only infect the marine benthic dinoflagellate Prorocentrum fukuyoi (Leander and Hoppenrath, 2008; Figure 1B). However, in 2017, Reñé et al. isolated a new species of Snorkelia from a different location (Catalan Coast, NW Mediterranean) infecting the dinoflagellate Levanderina fissa (Reñé et al., 2017a). The host range of Snorkelia species, therefore, remains enigmatic. For the “true” generalists, P. sinerae, P. corolla, and P. rostrata, the host range was determined mainly by in vitro cross-infection. In vitro experiments could result in artificial linkage between a host species and a parasite, which is not representative of the ecological reality (Råberg et al., 2014; Alacid et al., 2016; Reñé et al., 2021b). Infection of non-preferred hosts could be based on chemical and physiological processes shared by several related hosts (e.g., Garcés et al., 2013b). Furthermore, when a parasite infects a host but is not able to produce viable parasitic cells, such “host” cannot be considered as a “true” host for the parasite, e.g., infection of a chlorophyte strain of Pyramimonas by P. corolla (Rodríguez and Figueroa, 2020). The infection process is probably the result of plasmomorph mechanisms between P. corolla’s hosts, but the parasite cannot achieve its whole life cycle. In this review, we considered Parvilucifera species as generalists when they realize a viable life cycle in a large repertoire of host species. A BHR gives them the full potential of host shifting in new environments, although there are evidences of host preference in nature (Alacid et al., 2016). However, much remains to be done to evaluate the contribution of the plasmomorph and convergent traits of a parasite to the success of generalists.

Despite a “hot-spot” of detection in Europe, the distribution of these parasites of dinoflagellates is worldwide in marine environments (Figure 3B). As major contributors, with diatoms, to the fixation of inorganic carbon through photosynthesis (Falkowski et al., 1998), dinoflagellates are one of the most important components of marine phytoplankton as primary producers and grazers. Parvilucifera parasites could participate in regulating dinoflagellate populations in the blooming period, as shown in Spain where 5 to 18% of the population of the noxious dinoflagellates, Alexandrium minutum, were killed by Parvilucifera sp. during natural bloom (Alacid et al., 2017). For methodological limitation reasons, most studies describing Parvilucifera spp. have been carried out when prevalence is the highest on blooming host species, thus, increasing the chances to detect and describe new parasitic species. Unfortunately, this bias hides a possible important part of host and parasite diversity (Reñé et al., 2021b). Indeed, in some studies, the host range of these parasites appears to be much wider, including non-harmful species (e.g., Figueroa et al., 2008; Garcés et al., 2013a; Rodríguez and Figueroa, 2020), even if the in vitro specificity tests do not take into account the complexity of planktonic interaction network and the in situ parasitic host preferences. This lack of knowledge on Parvilucifera distribution, diversity, and trophic niche might become problematic, given its BHR nature and the global trades. Indeed, ballast waters, which are one of the most important vectors of introduction in marine systems (Carlton, 1985; Barry et al., 2008), could contribute to the dispersal of Parvilucifera trophonts or dormant sporangia (e.g., Darling et al., 2018), thus, increasing the risk of settlement in new areas. Hence, introductions of these BHR parasites could act positively as a top-down control of the toxic dinoflagellate blooming species and, conversely, could play a key role in deregulating the planktonic compartment given the importance of dinoflagellates as primary producers inducing significant shifts in marine food webs.

Xcellidae, a New Group of Fish Parasites

In 1969, Brooks et al. of two flatfish species, the black-tipped flounder, Psettichthys melanostictus, and the Pacific plaice, Platichthys stellatus (Brooks et al., 1969). Thirty-five years later, a protozoan named X-cell was identified as responsible for this pathology inducing also multiple lesions and swelling of the gill filaments (Miwa et al., 2004). Using small and large concatenated subunit (SSU and LSU) rDNA phylogeny reconstruction, Freeman et al. (2017) investigated the phylogenetic position of this enigmatic X-cell protist and showed that it formed two highly distinct clades, one corresponding to the pseudobranchial parasites of Gadiformes, and the other to gill and epidermal X-cells from Perciformes and Pleuronectiformes called Gadixcellia and Xcellia, respectively. More recently, Karlsbakk et al. (2021) described a new genus named Salmoxcellia, a sister group to Gadixcellia. Despite an unusually high genetic divergence between the three genera, with a similarity of 74.9% for the SSU rDNA sequences, these two close sister clades form a new family called Xcellidae (or Xcellins) branching within the Perkinsea lineage (Freeman et al., 2017; Karlsbakk et al., 2021) (Figure 2B).

Histological examination and scanning electron microscopy of xenomas of infected fish revealed a large number of clustered round parasitic cells surrounded by host connective tissue. Currently, only one developmental life stage has been described up to today; hence, the life cycle of this parasite still remains unknown (Freeman et al., 2017). However, Freeman et al. (2017) suggested that infection occurs via contact between fish and the benthos. Indeed, X-cell infection occurs in fish species with at least one benthic stage during their life cycle (Fahay, 1983; Lough et al., 1989). An experimental infection in situ
in tanks revealed that only fish exposed, even transiently, to the benthos were infected and showed symptoms of disease (Eydal et al., 2010). The infected fish samples are mainly young and adult individuals. Indeed, the first pathogenic symptoms (pseudobranchial xenomas) appear macroscopically in young wild cod (around 6 months old, 6- to 13-cm length) from Icelandic waters with a prevalence peak of 23% at the age of 22 months, whereas older or larger fish (40- to 76-cm length) have a lower prevalence around 7 and 1%, respectively (Eydal et al., 2010). Similar trends had been observed in Atlantic cod (Morrison et al., 1982) and the Pacific cod (Stich et al., 1976). However, the question remains open whether the mortality induced by the X-cell infectious agent on juveniles could remove an age class from the field sampling and bias the observed prevalence. Although mortality induced by these infectious agents still needs to be demonstrated, the transfer of wild infected individuals of Gadus morhua in tanks has shown an important mortality (Eydal et al., 2010).

Parasites belonging to these three genera have been detected in more than 20 fish species belonging to five orders of teleost: Pleuronectiformes (flatfishes) (Freeman, 2009; Freeman et al., 2011), Perciformes (perch-like fish) (Katsura et al., 1998), Gadiformes (cods) (Freeman et al., 2011), Siluriformes (catfishes) (Diamant et al., 1994), and Salmoniformes (salmonids) (Dyková et al., 1993; Karlsbakk et al., 2021). Until now, five species have been described (Figure 1B). The parasite Xcellia pleurocentri infection has been confirmed in Hippoglossoides dubius and Pseudopleuronectes obscurus but also suspected in many other species from the same area: Cleisthenes herzensteini, C. pinetorum, Glyptocephalus stelleri, Kareius bicoloratus, Hippoglossoides elassodon, Lepiseta pinnifasciata, Platichthys stellatus, Parophrys vetulus, Pseudopleuronectes schrenki, and Verasper moseri. Equally, Xcellia lamelliphila infection is detected in Limanda limanda, Lycodes spp., Macrouronus novaeezelandiae, Merluccius gayi gayi, and Trematomus spp. (Freeman et al., 2017). Finally, the newly described parasite Salmoxcellia vastator is detected in Oncorhynchus mykiss and in Salmo salar (Karlsbakk et al., 2021). These three parasites are, therefore, BHR parasites (Figure 1B). Conversely, Xcellia gobii and Gadixcellia gadi infecting, respectively, Acanthogobius flavimanus and Gadus morhua are NHR parasites and classified as specialists (Freeman et al., 2017; Figure 1B). This classification will certainly evolve as new knowledge on Xcellidae group becomes available.

In addition, Xcellidae protists have been detected in a restricted geographical area (Northern Europe and Japan) (Figure 3C), but the monitoring of these X-cell infections is hard to carry out and stay largely incomplete because of migratory fish hosts in the worldwide ocean. Their host range is mainly unknown, but we can hypothesize that it may be broader than that currently identified. An extended host range would not be surprising considering the Perkinsinea phylum constituted of a majority of BHR parasites closely related phylogenetically (e.g., the Perkinsidae) (Figure 2B). Overall, the consequences of the X-cell infection on fish populations are not well known. However, as it affects important halieutic resources as described for the salmonid parasite, Salmoxcellia vastator, it makes farmed salmonid fillets unsuitable for sale (Karlsbakk et al., 2021). Moreover, these parasites could infect vulnerable species, such as Atlantic cod fish G. morhua, extensively farmed, but mainly in the north of Europe (see FAO, 2021 Programme—Gadus morhua). Commercially important fish, a concern of the global trade, which could act as reservoirs allowing the dissemination in farmed and wild fish populations.

Severe Perkinsinea Infection, an Infectious Agent of Tadpoles Populations

With up to 50% of all threatened species, amphibian populations are emblematic representatives of the sixth mass extinction event to which highly virulent wildlife diseases contribute (Stuart et al., 2004; Gewin, 2008; Chambouvet et al., 2020). Recent works have mainly identified two pathogens as drivers of this decline, the fungus Batrachochytrium dendrobatidis and the Ranavirus (e.g., Bosch et al., 2001; Fisher et al., 2009; Gray et al., 2009; Olson et al., 2013).

However, in April 2006, in a pond in northeast Georgia (United States), Davis et al. (2007) observed a massive mortality event of southern leopard tadpoles, Rana sphenoecephala (syn. Lithobates sphenoecephalus), attributed to an unknown protist affiliated by phylogenetic analysis to the Perkinsinea lineage. Using phylogenetic analysis, it has been shown that this organism belongs to a discrete clade named pathogenic Perkinsinea clade (PPC) (Isidoro-Ayza et al., 2017) within the wider monophyleic cluster of the Novel Alveolate Group 01 (NAG01) (Chambouvet et al., 2015) obtained from disparate freshwater environments or internal organs (e.g., liver tissues) of a wide variety of Neobatrachia suborder (Chambouvet et al., 2015; Isidoro-Ayza et al., 2017; Figures 1B, 2B). Infectious agents of the PPC clade have been identified as responsible for the die-offs of tadpole throughout the United States (Isidoro-Ayza et al., 2017). Infected tadpoles with pathological symptoms, named “severe Perkinsinea infection” (SPI), showed lethargic swimming with enlarged and histopathologic lesions of the liver, mesonephros, spleen, pancreas, gills, gastrointestinal tract, skeletal muscle, dermis, and peritoneum (Green et al., 2002; Davis et al., 2007; Jones et al., 2012; Isidoro-Ayza et al., 2017). Histological examination of infected liver tissues revealed a massive number of round infective cells replacing the normal hepatic parenchyma. Two distinct putative life stages were identified invading the hepatocyte cells corresponding to hypnospore-like and trophozoite-like life stages (Isidoro-Ayza et al., 2017, 2019). SPI symptoms have been reported from summer to early autumn in boreal and temperate regions, and from late winter to early spring in subtropical areas (Isidoro-Ayza et al., 2017; Figure 3D). Although this parasite is mainly described as infecting tadpole life stage, one report highlights infection of the adult populations with granulomatous lesion in the legs (Jones et al., 2012). These results suggest that either most of the tadpoles die before the metamorphosis or that mature immune systems acquired after the metamorphosis may drive back the infection (Isidoro-Ayza et al., 2017). Although the relationship between the infectious agent and the disease is not yet well established, symptoms of the disease...
could result from co-infection between the Perkinsia parasite and others infectious agents, such as the FV3-like virus, responsible for frog population declines, which infects both larval and adult amphibians (Gray et al., 2009; Lesbarrères et al., 2012), and/or alteration of tadpole immune systems (Isidoro-Ayza et al., 2017).

During monitored SPI outbreaks, mortality rate can reach up to 95% of the population leading to the loss of an entire age class or, in case of chronic infection, to a reduced recruitment (Green et al., 2002; Isidoro-Ayza et al., 2017, 2019). Now recognized as the third most common infectious disease of anuran species, SPI infectious agent is described as an emerging pathogen. It is therefore now fundamental to understand the relationship between this infectious agent and the outcome of the disease and how other pathogen communities and/or environmental factors could affect disease susceptibility. Furthermore, as the globalization of trade where amphibian species, generally involved in meat or pet trade, could spread the parasitic invaders via reservoir species into native and/or naive amphibian populations, the study of this pathogen is crucial.

Environmental Diversity: Unveiling Putative Pathogens?

Molecular methodologies have recently highlighted that our vision of Perkinsia diversity, mainly assessed by culture-based methods, is still only the tip of the iceberg. In both marine and freshwater environments, analysis of genetic diversity based on SSU rDNA sequences of the smallest sizes of plankton (<5 μm), revealed a previously unknown diversity of these organisms (Moon-van der Staay et al., 2001; López-Garcia et al., 2001, 2003; Lefranc et al., 2005; Magot et al., 2009).

A recent overview of freshwater Perkinsia genetic diversity, especially in lakes, revealed that this group appears to be diverse and abundant, with an observation of 2,927 operational taxonomic units (OTUs) defined at 95% of sequence similarity, which represented 3% of the eukaryotic diversity analyzed in Debros et al. (2017), suggesting an important role in the trophic web (e.g., Lepère et al., 2008, 2011; Jobard et al., 2020). Recent studies have highlighted that freshwater clades could be a parasitic protist of the colonial green algae Sphaerocystis sp. (Jobard et al., 2020). These results are consistent with the description of Rastrimonas gen. nov. (Brugerolle, 2003), previously described as Cryptophagus subtilis, infecting the free-living Cryptophyte Chilomonas paramaccum (Brugerolle, 2002). However, apart from these two descriptive publications, no molecular analysis has yet been carried out to definitively affiliate Rastrimonas sp. within the Perkinsia lineage.

In marine environments, molecular signatures of Perkinsia were also found in extreme environments, such as hydrothermal vents or anoxic fjords (López-Garcia et al., 2003; Zuedendorf et al., 2006), but most of environmental genomics studies target water column samples [e.g., surface or deep chlorophyll maximum (DCM) depths] (e.g., Moon-van der Staay et al., 2001; López-Garcia et al., 2003; De Vargas et al., 2015). This lack of genetic signatures in marine environmental surveys represent a real paradox, since the two main cultivable groups described, Perkinsus spp. and Parvilucifera spp., are marine. In 2014, targeting the V4 hypervariable region of the rDNA and rRNA templates using 454 sequencing technology, Chambouvet et al. (2014) evaluated the genetic diversity of the eukaryotic microbial community in two sampling depths (surface and DCM) and in the sediment across four different locations across Europe (Oslo, Norway; Naples, Italy; Barcelona, Spain; Roscoff and France). The analysis revealed an unexpected genetic diversity of ribosomally active organisms belonging to Perkinsia other than than Parviluciferaceae and Perkinsus clusters mainly detected in the sediment with a total of 265 sequences clustered in 150 OTUs defined at 99% similarity. The ribosomal RNA sequences present in these clades belong to metabolically active organisms, which certainly play an active role in the ecosystem functioning. This cryptic diversity, only detected by their genetic signatures, raises new scientific questions about the putative existence of non-parasitic heterotrophic organisms, the host range of parasitic Perkinsia, and their role and impact on the aquatic food web. Finally, these results suggested that sediments can act as parasite reservoirs, as suggested for Amoebophrya parasites (Chambouvet et al., 2011) and for marine Perkinsia (Chambouvet et al., 2014; Freeman et al., 2017; René et al., 2021b). However, studies on diversity and distribution are scarce and restricted to specific environments (e.g., coastal water or few freshwater environments). A wider study of the Perkinsia distribution could give a more complete picture of environmental distribution unveiling potential reservoirs of pathogens. A phylogenetic analysis of Perkinsia illustrates two main evolutionary divisions between clades associated to Perkinsidae and Xcellidae, and Parviluciferaceae prior to marine/freshwater specialization (Jobard et al., 2020). This also suggests that transitions between marine and freshwater were few in the life history of Perkinsia, as differences between these environments constitute a barrier for cross-colonization for most organisms (Logares et al., 2009; Brête et al., 2010). However, a deep analysis of environmental DNA may reveal new information on the evolution and environmental colonization of Perkinsia.

CONCLUSION

All currently known members of Perkinsia are described as parasitic species infecting protists, mollusks and vertebrates. They are detected in all ecosystems from the tropics to high latitudes and from freshwater to marine environments (Brête et al., 2010; Magot et al., 2011). Perkinsia share common characteristics, which represent serious threats for biodiversity and human activity.

(1) They are putatively predominantly BHR. Perkinsia consist of four described lineages of putative BHR parasites, with strong evidences of host preference (e.g., Garcés et al., 2013a) or susceptibility (e.g., Calvo et al., 1999). At high taxonomic level, such homogeneity is rare compared with other major parasitic clades in ecosystems. Indeed, some strains of Amoebophrya species (Syndiniales, Alveolata) exhibit different specificity to host species...
leading to a mix of parasitic protist displaying various degrees of host specialization (Coats and Park, 2002; Chambouvet et al., 2008; Farhat et al., 2018). In the same way, while many chytrid fungi (Chytridiomycota) infecting phytoplankton are mainly considered as host specific (Ibelings et al., 2004), Dinomycetes arensensis has a BHR focused on dinoflagellate species (Lepeltier et al., 2014a). The same trend is observed for Haplosporidiae, which is one of the major pests involved in mollusks diseases with the Perkinsidae, which is composed of NHR and BHR. For example, Bonamia species infects several oyster species (Engelsma et al., 2014) whereas Haplosporidium species are able to infect distant phylogenetic hosts, including mollusks, crustaceans, or even polychaetes (Arzul and Carnegie, 2015). This peculiar putative BHR dominant characteristic is probably strongly related to their common evolutionary history and may represent an ecological advantage that enable survival improvement via a wide variety of “potentials” hosts. However, this assumption should be taken into caution as variability in pathogenicity could also rely on the existence of specific strains within a species. For example, Marteilia refringens (Paramyxin), responsible of marteliosis, exhibits a potential profile of BHR parasite because this microeukaryotic parasite infects oysters, mussels, and clams. Nonetheless, three different strains have been observed: the “O” strain preferentially found in oysters, the “M” strain mainly found in mussels, and finally the recent “C” strain infecting the cockles Cerastoderma edule (Le Roux et al., 2001; Carrasco et al., 2012; Arzul et al., 2014; Guo and Ford, 2016). Hence, the putative BHR of Perkinsidae, Parviluciferaceae, and SPI agents may finally represent the tip of the iceberg with a more complex reality that needs to be urgently explored because of consequences in terms of the development and implementation of new strategies in disease and conservation management.

(2) They are pathogenic for many keystone, engineer, or endangered species. Like other important parasites listed at international or national level (e.g., Batrachochytrium dendrobatidis or Marteilia refringens), Perkinsea clearly contributes to the loss and dismiss of some species (e.g., the SPI agent) and to the massive loss of fishery and aquaculture resources (e.g., P. olseni and P. marinus). Recent years have seen the advent of diseases induced by Batrachochytrium dendrobatidis and B. salamandrivorans (agent of chytridiomycosis), listed in the OIE-notifiable disease list, which are mainly responsible for the collapse of amphibian populations (Scheeke et al., 2019). In 2020, Chambouvet et al. reviewed new infectious diseases concerning Apicomplexans (Coccidians, Gregarines) and Perkinsea (SPI agent) that may worsen amphibian situation. Today, many species of frogs, such as Lithobates capito or L. sevosus, have been listed as “Near threatened” to “Critically endangered” status on the International Union of Conservation of Nature (IUCN) red list (Chambouvet et al., 2020). Their involvement in this loss may contribute to major shifts in biological communities.

(3) They are responsible for economical loss. Economic valuable species are affected by Perkinsea infection, which may lead to collapse of fishery and aquaculture industries. As protozoans caused most of the historically significant diseases in molluscs, this parasitic group is of major concern for the shellfish exploitation (Carnegie et al., 2016). The Dermo disease (etiological agent: P. marinus) is one of the major marine molluscan diseases in addition to MSX (multinucleated sphere unknown caused by Haplospordium nelsoni), marteliosis (Marteilia refringens and M. sydneyi), and bonamiosis (Bonamia ostreae and B. exitiosa) (Arzul and Carnegie, 2015; Guo and Ford, 2016). Globally, Haplosporidia and Perkinsidae are two major threats for mollusk health even if the mortality rate produced by P. marinus (max. 60%) on oysters is lower than the mortality rate of Haplosporidium nelsoni or Marteilia refringens (90–100%) on the same resource (Guo and Ford, 2016). Indeed, in the Chesapeake Bay, the oyster harvest decline (~80.000t between 1910 and 1980 to 15.000t in 1986) was attributed to both parasites P. marinus and H. nelsoni in a context of inadequate management practices (Héral et al., 1990; Goulleque et al., 1994). Recurrent declines in clam harvests are recorded as a result of massive mortalities caused by P. olseni in Korea (Park et al., 1999; Choi and Park, 2005), China (YuBo et al., 2001), and Japan (Hamaguchi et al., 1998). The Manila clam landings from culture in 1997 was approximately 14.000t, which is only one-fifth of the clam landings in 1990 (Park and Choi, 2001). In Europe (Spain, Portugal, and Italy), the parasite destroyed R. decussatus (native) and R. philippinarum (exotic) populations (Azevedo, 1989; Figueras et al., 1992; Pretto et al., 2014). Perkinosis induces more damage to clam stocks than other important diseases, such as the brown ring disease (BRD), producing approximately 20% of mortality over 2 years (Guo and Ford, 2016).

(4) They are easily translocated and could be invasive. Today, many precautions (e.g., quarantine of animals) must be taken around the world to stem the spread of Perkinsea, particularly with regard to Perkinsidae. Since the first mortalities, P. marinus has been listed as a notifiable pathogen by the OIE (Oie-Listed Diseases, 2021: OIE—World Organisation for Animal Health) and the European Commission (Directive 2006/088/EC). However, P. olseni is solely classified at the international level in the OIE—list of notifiable diseases but is out of concern for the European Commission (Carnegie et al., 2016). Some European countries are Perkinus olseni free, and its exclusion from the notifiable disease list from the European Commission can lead to relaxed vigilance within the trading network contributing to its spread in European non-affected areas (Carnegie et al., 2016). National surveillance efforts are different between European member states, and mortality events are mainly reported by shellfish farmers. However, these networks have already revealed pathogens in new areas, like OsHV-1 μvar
or B. ostreae, or new pathogens implied in shellfish mortalities, like Mikrocytos species. Therefore, given the putative wide host range of P. osleni and other Perkinsea species, the detection of new pathogens in new areas should be monitored using next-generation sequencing methodologies with confirmation by classical microscopic techniques.

(5) Most Perkinsea are only described by their genetic signatures.

Environmental sequencing revealed a high diversity of several potential microeukaryotic parasites in the water column and sediments (Chambouvet et al., 2014). These results show that the study of environmental diversity is absolutely crucial to identify the potential distribution and emergence of parasites. Thanks to molecular methods, some of problematical parasitic groups, e.g., the haplosporidians, are under survey (Hartikainen et al., 2014; Bass et al., 2015). Recently, Haplosporidium diporeiae infecting amphipods was associated to the previously described environmental clade “haplosporidian clade C” (Hartikainen et al., 2014; Winters and Faisal, 2014; Bass et al., 2015). It is now clear that Perkinsea lineage is genetically diverse in aquatic environments and may be composed of clades with a pathogenic potential whose hosts have not yet been identified. This environmental diversity described across different clusters (e.g., Chambouvet et al., 2014) clearly represents potential parasites. However, their host range and impact of the aquatic food webs are still black boxes that the scientific community need to urgently address considering the impact of already described BHR species belonging to this lineage.

In a context of intensification of the global trade, this opportunistic BHR parasitic protists represent a threat to become successful invasive species eventually leading to a new putative emerging disease. It is thereby now important to investigate their full host range, their cryptic diversity, and their role in the global aquatic network. On the other hand, a growing body of evidence emphasizes the importance of the interactive network between a host, the whole associated microbial communities (including others parasites), and the environmental conditions in the determination of infection outcome. Indeed, the outdated “one parasite, one disease” paradigm is not sufficient to explain a disease, and therefore, these parasites and their studies should be integrated into a larger scheme (Bass et al., 2019). Knowing under what conditions invasion by these infectious agents could be successful in and what threats they could pose to native host populations are two more fundamental questions that need to be addressed.

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SI, PS, and AC designed the study. SM performed the phylogenetic analyses. AC provided the funds. All authors drafted the manuscript and approved the final version of the manuscript.

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**SUPPLEMENTARY MATERIAL**

The Supplementary Material for this article can be found online at: https://www.frontiersin.org/articles/10.3389/fmicb.2021.735815/full#supplementary-material

**REFERENCES**

Agosta, S. J., Janz, N., and Brooks, D. R. (2010). How specialists can be generalists: resolving the “parasite paradox” and implications for emerging infectious disease. Zoologa 27, 151–162. doi: 10.1590/S1984-46702010000200001

Alacid, E., Park, M. G., Turon, M., Petrou, K., and Garcés, E. (2016). A game of russian roulette for a generalist dinoflagellate parasitoid: host susceptibility is the key to success. Front. Microbiol. 7:769. doi: 10.3389/fmicb.2016.00769

Alacid, E., René, A., Camp, J., and Garcés, E. (2017). In situ occurrence, prevalence and dynamics of Parvilucifera parasitoidae during recurrent blooms of the toxic dinoflagellate Alexandrium minutum. Front. Microbiol. 8:1624. doi: 10.3389/fmicb.2017.01624

Alacid, E., René, A., Gallisai, R., Palohéimo, A., Garcés, E., and Kremp, A. (2020). Description of two new coexisting parasitoids of blooming dinoflagellates in the Baltic sea: Parvilucifera catillosa sp. nov. and Parvilucifera sp. (Perkinsea, Alveolata). Harmful Algae 100:101944. doi: 10.1016/j.hal.2020.101944

Alacid, E., René, A., and Garcés, E. (2015). New insights into the parasitoid Parvilucifera sinerae life cycle: the development and kinetics of infection of a bloom-forming dinoflagellate host. Protist 166, 677–699. doi: 10.1016/j.protis.2015.09.001

Allam, B., Carden, W. E., Ward, J. E., Ralph, G., Winnicki, S., and Pales Espinosa, E. (2013). Early host-pathogen interactions in marine bivalves: evidence that the alveolate parasite Perkinsus marinus infects through the oyster mantle during rejection of pseudofeces. J. Invertebr. Pathol. 113, 26–34. doi: 10.1016/j.jip.2012.12.011

Anderson, R. S., Unger, M. A., and Burreson, E. M. (1996). Enhancement of Perkinsus marinus disease progression in TBT-exposed oysters (Crassostrea virginica). Mar. Environ. Res. 42, 177–180. doi: 10.1016/0141-1136(95)00066-6

Andrews, J. (2013). Early host-pathogen interactions in marine bivalves: evidence that the alveolate parasite Perkinsus marinus infects through the oyster mantle during rejection of pseudofeces. J. Invertebr. Pathol. 113, 26–34. doi: 10.1016/j.jip.2012.12.011

Anderson, R. S., Unger, M. A., and Burreson, E. M. (1996). Enhancement of Perkinsus marinus disease progression in TBT-exposed oysters (Crassostrea virginica). Mar. Environ. Res. 42, 177–180. doi: 10.1016/0141-1136(95)00066-6

Andrews, J. (1996). History of Perkinsus marinus, A pathogen of oysters in Chesapeake Bay 1930–1984. VIMS Articles 503:5.

Arzul, I., and Carnegie, R. B. (2015). New perspective on the haplosporidian Perkinsus marinus, A pathogen of oysters in Chesapeake Bay 1930–1984. VIMS Articles 503:5.

Available online at: https://www.frontiersin.org/articles/10.3389/fmicb.2021.735815/full#supplementary-material
Bråte, J., Logares, R., Berney, C., Ree, D. K., Klaveness, D., Jakobsen, K. S., et al. (1994). Growth and mortality of Eastern oysters, *Crassostrea virginica* (Apicomplexa, Perkinsea) parasite of the clam pathogen *Perkinsus marinus* under various environmental conditions. *Oceanogr. Lit. Rev.* 10.1007/s10551-015-9925-z

Burreson, E., Stokes, N., and Friedman, C. (2000). Increased virulence in an introduced pathogen: *Haplosporidium nelsoni* (MSX) in the eastern oyster *Crassostrea virginica*. *J. Aquat. Anim. Health* 12, 1-8.

Cáceres-Martínez, J., Ortega, M. G., Vásquez-Yeomans, R., García, T., de, J. P., Coen, L., and Grizzle, R. (2007). The importance of habitat created by molluscan shellfish to managed species along the Atlantic Coast of the United States. *Habitat. Manag. Ser.* 8, 1–108.

Carlo, J. T., and Geller, J. B. (1993). Ecological roulette: the global transport of nonindigenous marine organisms. *Science* 261, 78–82. doi: 10.1126/science.261.5117.78

Carlton, J. T. (1985). Transoceanic and interoceanic dispersal of coastal marine organisms: the biology of ballast water. *Oceanogr. Mar. Biol.* 23, 31–137.

Carnegie, R. B., Ford, S. E., Crockett, R. K., Kingsley-Smith, P. R., Bienlien, L. M., Sali, L. S., et al. (2021). A rapid phenotype change in the pathogen *Perkinsus marinus* was associated with a historically significant marine disease emergence in the eastern oyster. *Sci. Rep.* 11:12872. doi: 10.1038/s41598-021-92379-6

Carrasco, N., Andree, K. B., Lacuesta, B., Roque, A., Rodgers, C., and Furones, M. D. (2012). Molecular characterization of the *Maretiella* parasite infecting the common edible cockle *Cerastoderma edule* in the Spanish Mediterranean coast. *Aquaculture* 324–325, 20–26. doi: 10.1016/j.aquaculture.2011.10.017

Chambouvet, A., Alves-de-Souza, C., Cuffe, V., Marie, D., Karpov, S., and Guillou, L. (2011). Interplay between the parasite *Amoebophrya* sp.(Alveolata) and the cyst formation of the red tide dinoflagellate * Scrippsiella trochoidea*. *Protist* 162, 637–649.

Chambouvet, A., Berney, C., Romac, S., Audic, S., Maguire, F., De Vargas, C., et al. (2014). Diverse molecular signatures for ribosomally active *Perkinsea* in marine sediments. *MBM Microbiol. 14*:110. doi: 10.1186/1471-2180-14-110

Chambouvet, A., Gower, D. J., Jirků, M., Yabsley, M. J., Davis, A. K., Leonard, G., et al. (2015). Cryptic infection of a broad taxonomic and geographic diversity of tadpoles by *Perkinsea protozoa*. *Protist* 166(1), 55–70.

Chintala, M., Bushek, D., and Ford, S. (2002). Comparison of *in vitro*-cultured and wild-type *Perkinsus marinus*. II. Dosing methods and host response. *Dis. Aquat. Org.* 51, 203–216. doi: 10.1354/dao051203

Choi, K.-S., and Park, K.-I. (2005). *Current Status of Perkinsus Infection in Korean Waters Diseases in Asian Aquaculture*. V. Manila: Fish Health Section, Asian Fisheries Society. 263–274.

Choi, K.-S., and Park, K.-I. (2010). "Review on the protozoan parasite *Perkinsus olseni* (Lester and Davis, 1981) infection in Asian waters," in Coastal Environmental and Ecosystem Issues of the East China Sea, eds A. Isimatsu and H. J. Lie (Nagasaki: Nagasaki University), 269–281.

Coen, L., and Grizzle, R. (2007). The importance of habitat created by molluscan shellfish to managed species along the Atlantic Coast of the United States. *Habitat. Manag. Ser.* 8, 1–108.

Cohen, A. N. (1998). Accelerating invasion rate in a highly invaded estuary. *Science* 279, 555–558. doi: 10.1126/science.279.5350.555

Coss, A. C., Robledo, J. A. F., Ruiz, G. M., and Vasta, G. R. (2001). Description of *Perkinsus andrewsi* n. sp. isolated from the Baltic clam (*Macoma balthica*) by characterization of the ribosomal RNA locus, and development of a species-specific PCR-based diagnostic assay. *J. Eukaryot. Microbiol.* 48, 52–61. doi: 10.1111/j.1550-7408.2001.tb00415.x

Cremonte, F., Balseiro, P., and Figueras, A. (2005). Occurrence of *Perkinsus olseni* (Protozoa: Apicomplexa) and other parasites in the venerid commercial clam *Pitar rostrata* from Uruguay, southwestern Atlantic coast. *Dis. Aquat. Org.* 64, 85–90. doi: 10.3354/dao064085
da Silva, P. M., Scardua, M. P., Vianna, R. T., Mendonça, R. C., Vieira, C. B., Dungan, C. F., et al. (2014). Two Perkinsus spp. infect Crassostrea gary oysters from cultured and wild populations of the Rio São Francisco estuary, Sergipe, northeastern Brazil. J. Invertebr. Pathol. 119, 62–71. doi: 10.1016/j.jip.2014.04.005

da Silva, P. M., Vianna, R. T., Guertler, C., Ferreira, L. P., Santana, L. N., Fernández-Bobo, S., et al. (2013). First report of the protozoan parasite Perkinsus marinus in South America, infecting mangrove oysters Crassostrea rhizophorae from the Paraíba River (NE, Brazil). J. Invertebr. Pathol. 113, 96–103. doi: 10.1016/j.jip.2013.02.002

Dairain, A., Legeay, A., and de Montaudouin, X. (2019). Influence of parasitism on bioturbation: from host to ecosystem functioning. Mar. Ecol. Prog. Ser. 619, 201–214. doi: 10.3354/meps12967

Dantas Neto, M. P., Gesteira, T. C. V., Sabry, R. C., Feijó, R. G., Forte, J. M., Boës, G., et al. (2016). First record of Perkinsus cheaspeakesi infecting Crassostrea rhizophorae in South America. J. Invertebr. Pathol. 141, 53–56. doi: 10.1016/j.jip.2016.10.007

Darling, I. A., Martinson, J., Gong, Y., Okum, S., Pilgrim, E., Lohan, K. M. P., et al. (2018). Ballast water exchange and invasion risk posed by intracoastal vessel traffic: an evaluation using high throughput sequencing. Environ. Sci. Technol. 52, 9926–9936. doi: 10.1021/acs.est.8b02108

Daszak, P. (2000). Emerging infectious diseases of wildlife – Threats to biodiversity and human health. Science 287, 443–449. doi: 10.1126/science.287.5452.443

Davis, A. K., Yabsley, M. J., Kevin Keel, M., and Maerz, J. C. (2007). Discovery of the North American amphibian ranavirus Parvulucfena sinerae sp. nov. (Perkinsozoa). Protist 159, 563–578. doi: 10.1016/j.protis.2008.05.003

Fisher, M. C., Garner, T. W. J., and Walker, S. F. (2009). Global emergence of Batrachochytrium dendrobatidis and amphibian chytridiomycosis in space, time, and host. Annu. Rev. Microbiol. 63, 291–310. doi: 10.1146/annurev.micro.091208.070115

Ford, S. E. (1996). Range extension by the oyster parasite Perkinsus marinus into the northeastern United States: response to climate change? Oceanogr. Lit. Rev. 12:1265.

Ford, S. E., and Smolowitz, R. (2007). Infection dynamics of an oyster parasite in its newly expanded range. Mar. Biol. 151, 119–133. doi: 10.1007/s00227-006-0454-6

Freeman, M. A., Eydal, M., Yoshimizu, M., Watanabe, K., Miura, K., and Legendre, P. (2017). Overview of freshwater microbial eukaryotes diversity: a first analysis of publicly available metabarcoding data. FEMS Microbiol. Ecol. 93:fix023. doi: 10.1093/femsme/fix023

Freeman, M. A., Fuss, J., Kristmundsson, Á, Bjorbækmo, M. F. M., Mangot, J.-C., and de Montaudouin, X. (2014). First record of Perkinsus chesapeaki infecting Crassostrea gasar in South America, infecting mangrove oysters Crassostrea rhizophorae from the Paraíba River, Sergipe, northeastern Brazil. Available online at: http://www.fao.org/fishery/culturedspecies/Gadusmorhua/en (accessed September 8, 2021).

Farihi, S., Florent, I., Noel, B., Kayal, E., Da Silva, C., Bigeer, E., et al. (2018). Comparative time-scale gene expression analysis highlights the infection processes of two Amoebophrya strains. Front. Microbiol. 9:2251. doi: 10.3389/fmicb.2018.02251

Figueras, A., Fernández Robledo, J., and Novoa, B. (1992). Occurrence of haplosporidian and Perkinsus-like infections in carpel-shell clams, Ruthigites decussatus (Linnaeus 1758), of the Rio de Vigo (Galicia, NW Spain). J. Shell. Res. 11, 377–382.

Figuerola, R. I., García, E., Massana, R., and Camp, J. (2008). Description, host-specificity, and strain selectivity of the dinoflagellate parasite Parvulucfena sinerae sp. nov. (Perkinsozoa). Protist 159, 563–578. doi: 10.1016/j.protis.2008.05.003

Freeman, M. A., Fuss, J., Kristmundsson, Á, Bjorbækmo, M. F. M., Mangot, J.-F., del Campo, J., et al. (2017). X-cells are globally distributed, genetically divergent fish parasites related to Perkinsids and Dinoflagellates. Annu. Rev. Microbiol. 71, 99–119. doi: 10.1146/annurev.micro.091208.070115

Freeman, M. A., Eydal, M., Yoshimizu, M., Watanabe, K., Miura, K., and Ogawa, K. (2011). Molecular identification and transmission studies of X-cell parasites from Atlantic cod Gadus morhua (Gadiformes: Gadidae) and the northern flounder Pseudopleuronectes obscurus (Pleuronectiformes: Pleuronectidae). Parasit. Vectors 4:15.

Freeman, M. A., Fuss, J., Kristmundsson, Á, Bjorbækmo, M. F. M., Mangot, J.-F., del Campo, J., et al. (2017). X-cells are globally distributed, genetically divergent fish parasites related to Perkinsids and Dinoflagellates. Curr. Biol. 27, 1645–1651.e3. doi: 10.1016/j.cub.2017.04.045

Gall, R. S. (2000). A sea under siege – alien species in the Mediterranean. Biol. Inv. 2, 177–186. doi: 10.1023/A:1001005701476

Gallard, D. (2014). Effects of bivalve aquaculture on the environment and their possible mitigation: a review. Fish. Aquac. J. 5:1–105. doi: 10.4172/2150-3508.1000105

Garcés, E., Alacid, E., Reñé, A., Petrou, K., and Simó, R. (2013b). Host-released dimethyl sulphide activates the dinoflagellate parasitoid Parvulucfena sinerae. ISME J. 7, 1055–1066. doi: 10.1038/isme.2012.173

Garcés, E., Alacid, E., Bravo, I., Fraga, S., and Figueroa, R. I. (2013a). Parvulucfena sinerae (Alveolata, Myzozoa) is a generalist parasitoid of dinoflagellates. Protist 164, 245–260. doi: 10.1016/j.protis.2012.11.004

Gewin, V. (2008). Riders of a modern-day Ark. PLoS Biol. 6:24. doi: 10.1371/journal.pbio.0060024

Goedken, M., Morsey, B., Sunila, I., Dungan, C., and Guise, S. D. (2009). The effects of temperature and salinity on apoptosis of Crassostrea virginica hemocytes and Perkinsus marinus. J. Shell. Res. 24, 177–183.

Goggin, L. (1992). Occurrence of parasites of the genus Perkinsus in France. Bull. Eur. Ass. Fish. Pathol. 12:174.

Gouletquer, P., Hérail, M., and Rotschild, B. (1994). Causes of decline of oyster production (Crassostrea virginica) in the Maryland portion of the Chesapeake Bay: a literature study. Haitiola 23, 87–112.

Gray, M., Miller, D., and Hovenman, J. (2009). Ecology and pathology of amphibian ranaviruses. Dis. Aquat. Org. 87, 243–266. doi: 10.3354/dao02138
Mangot, J.-F., Lepère, C., Bouvier, C., Debros, D., and Domaizon, I. (2009). Community structure and dynamics of small eukaryotes targeted by new oligonucleotide probes: new insight into the lacustrine microbial food web. Appl. Environ. Microbiol. 75, 6373–6381. doi: 10.1128/AEM.00607-09

McCauley, D. J., Pinsky, M. L., Palumbi, S. R., Estes, J. A., Joyce, F. H., and Warner, R. R. (2015). Marine defaunation: animal loss in the global ocean. Science 347:1255641. doi: 10.1126/science.1255641

Miwa, S., Nakayasu, C., Kamaishi, T., and Yoshiura, Y. (2004). X-cells in fish pseudotumors are parasitic protozoans. Dis. Aquat. Org. 58, 165–170. doi: 10.3354/dao058165

Moon-van der Staay, S. Y., De Wachter, R., and Vaulot, D. (2001). Oceanic 18S rDNA sequences from picoplankton reveal unsuspected eukaryotic diversity. Nature 409, 607–610. doi: 10.1038/35054541

Morley, N. J. (2010). Interactive effects of infectious diseases and pollution in aquatic molluscs. Aquat. Toxicol. 96, 27–36. doi: 10.1016/j.aquatox.2009.09.017

Morrison, C. M., Shum, G., Appy, R. G., Odensen, P., and Annand, C. (1982). Histology and prevalence of X-cell lesions in Atlantic cod (Gadus morhua). Can. J. Fish. Aquat. Sci. 39, 1519–1530. doi: 10.1139/f82-204

Norén, F., Moestrup, Ø, and Rehnstam-Holm, A.-S. (1999). Histology and prevalence of X-cell lesions in Atlantic cod (Gadus morhua). Can. J. Fish. Aquat. Sci. 39, 1519–1530. doi: 10.1139/f82-204

Park, K.-I., Choi, K.-S., and Choi, J.-W. (1999). Epizootiology of Perkinsus olseni in the wild and farmed Pearl Oyster, Pinctada fucata martensii. J. Shellfish Res. 18, 305–332. doi: 10.1080/35040695

Saikai, A. K., Allendorf, F. W., Holt, J. S., Lodge, D. M., Molofsky, J., With, K. A., et al. (2001). The population biology of invasive species. Annu. Rev. Ecol. Evol. Syst. 32, 305–332. doi: 10.1146/annurev.ecolsys.32.081501.114037

Saldarriaga, J. F., McEwan, M. L., Fast, N. M., Taylor, F. J. R., and Keeling, P. J. (2003). Multiple protein phosphogenes show that Oxysrrhis marina and Perkinsus marinus are early branches of the dinoflagellate lineage. Int. J. Syst. Evol. Microbiol. 53, 355–365. doi: 10.1099/ijs.0.02328-0

Sanil, N. K., Vijayan, K. K., Kripa, V., and Mohamed, K. S. (2010). Occurrence of the protozoan parasite, Perkinsus olseni in the wild and farmed Pearl Oyster, Pinctada fucata (Gould) from the Southeast coast of India. Aquaculture 299, 8–14. doi: 10.1016/j.aquaculture.2009.12.007

Seebens, H., Blackburn, T. M., Dyer, E. E., Genovesi, P., Hulme, P. E., Jeschke, J. M., et al. (2019). Amphibian fungal panzootic causes catastrophic and ongoing loss of biodiversity. Science 363, 1459–1463. doi: 10.1126/science.aaw379

Semmler, A., Ferreira, I., Timoneda, N., Sampedro, N., Alacid, E., Gallisai, R., Gordi, J., et al. (2015). Host preferences of coexisting Perkinsus parasites during coastal dinoflagellate blooms. Mol. Ecol. 30, 2417–2433. doi: 10.1111/mec.13589

Reece, K. (2000). Molecular epizootiology of Perkinsus marinus and P. chesapeaki infections among wild oysters and clams in Chesapeake Bay, USA. Dis. Aquat. Org. 82, 237–248. doi: 10.3354/dao01997

Reece, K. S., Siddall, M. E., Burreson, E. M., and Graves, J. E. (1997). Phylogenetic analysis of Perkinsus based on actin gene sequences. J. Parasitol. 83, 417–423.

Reñé, A., Alacid, E., Sampedro, N., Alacid, E., Gallisai, R., Gordi, J., et al. (2021). Host preferences of coexisting Perkinsus parasites during coastal dinoflagellate blooms. Mol. Ecol. 30, 2417–2433. doi: 10.1111/mec.13589

Reñé, A., Timoneda, N., Sampedro, N., Alacid, E., Gallisai, R., Gordi, J., et al. (2021b). Host preferences of coexisting Perkinsus parasites during coastal dinoflagellate blooms. Mol. Ecol. 30, 2417–2433. doi: 10.1111/mec.13589

Ruano, F., Batista, F. M., and Arcangeli, G. (2015). Perkinsos in the clams Ruditapes decussatus and R. philippinarum in the northeastern Atlantic and Mediterranean Sea: a review. J. Invertebr. Pathol. 124, 12–29. doi: 10.1016/j.jp.2015.07.015

Ruano, F., and Cachola, R. (1986). “Outbreak of a severe epizootic of Perkinsus marinus (LEVIN-87) at Ria de Faro clam’s culture beds,” in Proceedings of the 2nd International Colloquium on Pathology in Marine Aquaculture (Porto, Portugal), 41–43.

Ruíz, G. M., Rawlings, T. K., Dobbs, F. C., Drake, L. A., Mullady, T., Huq, A., et al. (2000). Global spread of microorganisms by ships. Nature 408, 49–50. doi: 10.1038/35040695

Ruíz, G. M., Rawlings, T. K., Dobbs, F. C., Drake, L. A., Mullady, T., Huq, A., et al. (2000). Global spread of microorganisms by ships. Nature 408, 49–50. doi: 10.1038/35040695

Siddall, M. E., Reece, K. S., Graves, J. E., and Burreson, E. M. (1997). “Total evidence” refutes the inclusion of Perkinsus species in the phylum Apicomplexa. Parasitology 115, 165–176. doi: 10.1017/S0031182097001157

Sinni, K., Vujanovic, J., Kripa, V., and Mohamed, K. S. (2010). Occurrence of the protozoan parasite, Perkinsus marinus in the wild and farmed Pearl Oyster, Pinctada fucata (Gould) from the Southeast coast of India. Aquaculture 299, 8–14. doi: 10.1016/j.aquaculture.2009.12.007

Skeehan, R. C., Pasmans, F., Skerratt, I. F. F., Berger, L., Martel, A., Beuken, W., et al. (2019). Amphilibian fungal panzootic causes catastrophic and ongoing loss of biodiversity. Science 363, 1459–1463. doi: 10.1126/science.aaw379

Thomas, F., and Poulin, R. (1998). Manipulation of a mollusc by a tropically transmitted parasite: convergent evolution or phylogenetic inheritance? Parasitology 116, 431–436. doi: 10.1017/S003118209802399X

Stich, H. F., Acton, A. B., and Forrester, C. R. (1976). Fish tumors and sublethal effects of pollutants. J. Fish Res. Board Can. 33, 1993–2001.

Stuart, S. N., Chanson, J. S., Cox, N. A., Young, B. E., Rodrigues, A. S. L., Fischman, D. L., et al. (2008). Status and trends of amphibian declines and extinctions worldwide. Science 306, 1873–1876. doi: 10.1126/science.1103538

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Valiulis, G. A., and Mackin, J. G. (1969). Formation of sporangia and zoospores by *Labyrinthomyxa* sp. parasitic in the clam *Macoma balthica*. *J. Invertebr. Pathol.* 14, 268–270. doi: 10.1016/0022-2011(69)90115-3

Van den Wyngaert, S., Goell, A. S., Spaak, P., and Ibelings, B. W. (2013). Herbicides in the environment alter infection dynamics in a microbial host-parasite system. Infection dynamics and environmental pollution. *Environ. Microbiol.* 15, 837–847. doi: 10.1111/j.1462-2920.2012.02874.x

Vilas, R., Cao, A., Pardo, B. G., Fernández, S., Villalba, A., and Martínez, P. (2011). Very low microsatellite polymorphism and large heterozygote deficits suggest founder effects and cryptic structure in the parasite *Perkinsus olseni*. *Infect. Genet. Evol.* 11, 904–911. doi: 10.1016/j.meegid.2011.02.015

Wilela, H. (1951). *Sporozoaires Parasites de la Palourde, Tapes decussatus* (L.). Lisboa: Revisita da Faculdade de Ciências, 379–386.

Wang, Y., Yoshinaga, T., and Itoh, N. (2018). New insights into the entrance of *Perkinsus olseni* in the *Manila clam*, *Ruditapes philippinarum*. *J. Invertebr. Pathol.* 153, 117–121. doi: 10.1016/j.jip.2018.03.005

Williamson, M., and Fitter, A. (1996). The varying success of invaders. *Ecology* 77, 1661–1666. doi: 10.2307/2265769

Winters, A. D., and Faisal, M. (2014). Molecular and ultrastructural characterization of *Haplosporidium diporeiae* n. sp., a parasite of *Diporeia* sp. (Amphipoda, Gammaridea) in the Laurentian Great Lakes (USA). *Parasit. Vectors* 7:343. doi: 10.1186/1756-3305-7-343

YuBo, L., XiChang, Z., LiJun, W., Bo, Y., Ying, Z., and ChunLei, C. (2001). Prevalence of *Perkinsus* sp. in the *Manila clam* *Ruditapes philippinarum* along northern coast of Yellow sea in China. *Oceanol. Limnol. Sin.* 32, 502–511.

Zuendorf, A., Bunge, J., Behnke, A., Barger, K. J.-A., and Stoeck, T. (2006). Diversity estimates of microeukaryotes below the chemocline of the anoxic Mariager Fjord, Denmark: diversity estimates of microeukaryotes. *FEMS Microbiol. Ecol.* 58, 476–491. doi: 10.1111/j.1574-6941.2006.00171.x

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