Adsorption of Hexavalent Chromium Using Banana Pseudostem Biochar and Its Mechanism

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Abstract: A low-cost biochar was prepared through slow pyrolysis of banana pseudostem biowaste at different temperatures, and characterized by surface area and porosity analysis, scanning electron microscopy (SEM), Fourier-transform infrared (FTIR) spectroscopy, and X-ray photoelectron spectroscopy (XPS). It was shown that the biochar prepared at low pyrolysis temperature was rich in oxygen-containing groups on the surface. Adsorption experiments revealed that the biochar prepared at 300 °C (BB300) was the best adsorbent for Cr(VI) with 125.44 mg/g maximum adsorption capacity at pH 2 and 25 °C. All the adsorption processes were well described by pseudo-second-order and Langmuir models, indicating a monolayer chemiadsorption. Furthermore, it was demonstrated that adsorption of Cr(VI) was mainly attributed to reduction of Cr(VI) to Cr(III) followed by ion exchange and complexation with the biochar.

Keywords: biochar; banana pseudostem; chromium; adsorption mechanism

1. Introduction

Chromium, a toxic heavy metal, is widely discharged into the aquatic environment from various industries [1,2], causing serious harm to human health directly or indirectly [3–5]. Usually, chromium exists in hexavalent and trivalent forms in polluted water. Hexavalent Cr is highly toxic, carcinogenic and allergenic, while trivalent Cr is an essential trace element for mammals and less toxic [2,6–8]. Many treatment methods have been developed for removal of hexavalent chromium from water in the past decades, such as chemical precipitation [9], sedimentation [10], flocculation [11], adsorption [12], ultrafiltration [13], ion exchange [14], chemical coagulation [15], and so on. Being effective and operationally simple, adsorption is a useful method for removal of hexavalent chromium, solving the problems of sludge disposal [16,17].

Biochar, a cost-effective and green carbon material prepared from biomass through thermal conversion in an oxygen deficient environment [18,19], has been widely used to remove or immobilize Cr(III) and Cr(VI) due to their large specific surface area, high porosity and abundant functional groups on surface [20,21]. As the adsorption capabilities of biochar are significantly affected by biomass types [20,21], a series of biomass including oak wood, oak bark [22], ramie [23], cotton stalk [24], beet tailing [25], coconut coir [26], leaf of Leersia hexandra Swartz [27] and so on have been employed to make biochars by pyrolysis at different temperatures for removal of hexavalent chromium in the past decades. The adsorption capability of these biochars ranged from 3.03 mg/g to 349.81 mg/g, depending on the biomass feedstock. However, the application of biochars for removal of Cr(VI) was limited by either low adsorption efficiencies or less production of biomass feedstocks. For example, the adsorption capability of Leersia hexandra Swartz leaf biochar was up to 349.81 mg/g, but the production of Leersia hexandra Swartz leaf was less; even oak wood was available in large quantities,
the adsorption capability of oak wood biochar was only 3.03 mg/g. As a result, biochars with high Cr(VI) adsorption capability from large-scale and low-cost biowaste are still highly necessary for remediation of Cr(VI) containing waste water.

Banana pseudostem, a major residual waste of banana which is one of most widely distributed consumed food crops, is massively generated every year [28]. Banana pseudostem consists mainly of cellulose, hemicellulose, pectin and lignin [29]. Here, banana pseudostem was used as material to prepare biochar through direct pyrolysis at different temperatures and both absorption capability and adsorption mechanism of the biochar for Cr(VI) were investigated in aqueous solution.

2. Materials and Methods

2.1. Materials

All chemicals including HCl, KOH, H₂SO₄, H₃PO₄, CH₃COCH₃, C₁₃H₁₄N₄O and K₂Cr₂O₇ were analytical grade reagents and purchased from Aladdin or Macklin in Shanghai of China. A stock solution (1000 mg/L) of Cr(VI) was prepared by dissolving K₂Cr₂O₇ in deionized water (Water treatment system, LD-UPW-20, Leading, Shanghai, China). The stock solution was then diluted to desired concentrations. The pH of Cr(VI) solution was adjusted by suitable concentration solution of HCl or KOH.

2.2. Preparation of Banana Pseudostem Biochars

Banana pseudostem, the layered pseudostem wrapped in the trunk of banana trees, was supplied by the farms around Haikou City, Hainan Province, China. The sample was air-dried at 60 °C by blast drying oven and then smashed to pass through a 60 mesh sieve (0.3 mm). The banana pseudostem powder was heated with a rate of 5 °C min⁻¹ until it reached target temperature (200, 300, 400, 500, 600 °C) and then maintained the temperature for 1 h in a muffle furnace with tubular reactor under suitable nitrogen flow rate. The resulting biochars were cooled down to room temperature under nitrogen flow. Then the biochar was washed with deionized water for several times, dried at 60 °C for 24 h. The dried biochars were stored in an airtight desiccator prior to use and were abbreviated as BB200, BB300, BB400, BB500 and BB600 respectively, according to the pyrolysis temperature.

2.3. Characterization of Banana Pseudostem Biochars

The textural property was analyzed using N₂ sorption at 77 K on an ASAP 2460 surface area and porosity analyzer (Micromeritics, Norcross, GA, USA). The biochars (0.1–0.2 g) were degassed at 100 °C in vacuum before test. The surface area, pore volume and average pore size were all calculated by the Brunauer-Emmett-Teller (BET) method. The microscopic properties and surface morphologies were characterized using a S-3000N scanning electron microscope (Hitachi, Tokyo, Japan) operated at 10 kV. The surface functional groups of the biochars were characterized by Fourier-transform infrared (FTIR) spectrometer (Bruker Tensor 27, Ettlingen, Germany). The biochars were grinded with KBr and rolled into sheets, then recorded between 400 and 4000 cm⁻¹. The surface elemental content (C, O, N and Cr) and surface functional groups were determined by X-ray photo-electron spectra (XPS) with an Escalab 250Xi spectrometer (Thermo Scientific, Waltham, MA, USA).

2.4. Adsorption Experiments

Batch adsorption experiments of Cr(VI) were performed by adding biochars (50 mg) and aqueous K₂CrO₄ solution (50 mL) into 100 mL sealed conical flasks and shaking at 180 rpm on a thermostat shaker for desired time. For effect of pH on adsorption efficiency, the adsorptions were tested at pH ranged from 1.0 to 5.0 using a 200 mg/L Cr(VI) concentration. For kinetic experiments, BB300, BB400, BB500 and BB600 were mixed with Cr(VI) solution at pH 2, respectively, then the mixtures were shaken for different time interval (0.5, 1, 2, 3, 5, 9, 13, 18, 24, 30, 36, 42 and 48 h). Adsorption isotherms were
carried out for 48 h by maintaining the temperature at 25 °C using different Cr(VI) concentrations (25, 50, 100, 200, 300, 400 and 600 mg/L) with initial pH 2.

The Cr(VI) concentrations before and after adsorption were determined on a UV-vis spectrophotometer (MAPADA UV-3300PC, Shanghai, China) at wavelength of 540 nm using 1.5-diphenylcarbazide as indicator. The adsorption amount of Cr onto banana pseudostem biochars was calculated by the difference of concentrations before and after adsorption. The adsorption capacities of biochars towards Cr(VI) were determined as follows:

\[ q_e = \frac{(C_0 - C_e)V}{M} \]  

where \( q_e \) (mg/g) is the adsorption capabilities of biochars; \( C_0 \) (mg/L) and \( C_e \) (mg/L) are the Cr(VI) concentrations before and after adsorption, respectively; \( V \) (L) is the adsorbate solution volume; \( M \) (g) is the dosage of biochar.

3. Results and Discussion

3.1. Characterization of Banana Pseudostem Biochars

The FTIR spectra of banana pseudostem biochars are shown in Figure 1. The peak at 3448 cm\(^{-1}\) was assigned to hydroxyl group, and the peaks at about 2928 and 2860 cm\(^{-1}\) were assigned to aliphatic –CH and –CH\(_2\) stretching vibrations, respectively [30]. The peaks at 1630 cm\(^{-1}\) resulted from both the C=O stretching vibration and C=C stretching vibration of aromatic ring. The peak at about 1433 cm\(^{-1}\) also corresponded to aromatic C=C stretching vibration of aromatic ring [31]. Obviously, the peak at 1630 cm\(^{-1}\) reduced and the peak at 1433 cm\(^{-1}\) increased while increasing the pyrolysis temperature to 500 and 600 °C, as decarbonylation and aromatization intensively took place at high temperature [32]. The peak at 1385 cm\(^{-1}\) was symmetrical stretching vibration of COO\(^-\) [33,34]. The peak at 1317 cm\(^{-1}\) was assigned to aliphatic CH\(_2\) deformation vibration [31]. The peaks at 1117 cm\(^{-1}\) and 1067 cm\(^{-1}\) represented alkoxo C–O and aromatic C–O stretching vibrations, respectively. The shift of C–O peaks also confirmed that aromatization took place as increasing pyrolysis temperature [35]. The peaks at 876 cm\(^{-1}\) and 781 cm\(^{-1}\) were assigned to Si–O–Si symmetric stretching vibrations [36].

![Figure 1. The Fourier-transform infrared (FTIR) spectra of banana pseudostem biochars.](image)

The surface area, pore volume and average pore size of banana pseudostem biochars are given in Table 1. The biochars were of mesoporous structure [37], the average pore diameter (PD) and pore volume (PV) increased with pyrolysis temperature. The BET surface area (SA) slightly increased as
The C 1s XPS peaks of banana pseudostem biochars are shown in Figure 2, and the ratios of C–O, C=O, COO and CO$_3^{2-}$ bonds to the C–C bonds based on peak area are presented in Table 2. The total content of O-containing groups (TCOG) including C–O, C=O, COO and CO$_3^{2-}$, an important parameter affecting the adsorption capability of biochars [40], generally decreased as pyrolysis temperature increased from 300 to 600 °C, as previously reported by Li [35]. Both C–O and C=O bonds sharply decreased, and COO and CO$_3^{2-}$ basically remained unchanged, as the pyrolysis temperature increased from 300 to 400 °C. This might be due to the intense decomposition of cellulose and hemicellulose by breaking the C–O and C=O groups [41–43]. While the pyrolysis temperature increased to 500 °C, the C=O group sharply reduced with slight reduction of CO$_3^{2-}$, and the COO obviously increased, indicating C=O and CO$_3^{2-}$ groups generated at low temperature were broken and COO groups were created at high temperature. With further increase of pyrolysis temperature to 600 °C, there were not significant changes of O-containing groups, as the degradation became relatively gradual when the temperature was up to 500 °C [26,32].

Figure 2. The XPS C 1s scans and peak fitting for banana pseudostem biochars.
While the pyrolysis temperature was further raised to 600 °C, which was favourable for the electron interaction between the biochars and chromate anions. Moreover, the oxidizing ability of chromate was significantly improved at low pH, leading to more Cr(VI) adsorption. For example, the adsorption of BB300 increased gradually from 6.94 mg/g to 18.36 mg/g with the decrease of pH from 5.0 to 3.0, then increased sharply to 189.89 mg/g as the pH decreased to 1.0. This was caused by the formation of massive positive charges on biochar surface and $\text{HCrO}_2^-$ in the solution, and the maximum adsorption capacity was obtained at pH 1, which fully indicated that the biochars were more active under acidic conditions for Cr(VI) adsorption. For example, the adsorption efficiencies of all banana pseudostem biochars generally increased with decreasing initial pH of Cr(VI) solution, and the maximum adsorption capacity was obtained at pH 1, which fully indicated that the biochars were more active under acidic conditions for Cr(VI) adsorption.

The surface morphologies of biochars are shown in Figure 3. The SEM image of BB300 revealed rod bundles with plicated surface and irregular small pieces, and BB400 also showed similar morphology. The SEM image of BB500 indicated that the rod bundles were partly destructed and became incompact. While the pyrolysis temperature was further raised to 600 °C, the rod bundles were almost destructed to small pieces and irregular sheets with some small strips.

**Table 2.** The peak area ratios of C–O, C=O, O–C=O and CO$_2^-$ bonds to the C–C bonds of C 1s XPS spectra of biochars.

| B.E. (eV) Ass. | BB300 | BB400 | BB500 | BB600 | BB300+Cr |
|---------------|-------|-------|-------|-------|----------|
| C1 (286.0) C–O (%)  | 34.14 | 26.25 | 26.94 | 29.56 | 43.52    |
| C2 (287.8) C=O (%)  | 12.62 | 8.50  | 0.11  | 0.13  | 0.52     |
| C3 (288.5) COO (%)  | 0.74  | 0.85  | 5.20  | 5.73  | 8.81     |
| C4 (289.0) CO$_2^-$ (%) | 10.47 | 10.39 | 9.85  | 7.19  | 5.35     |
| TCOG            | 57.97 | 45.99 | 42.10 | 42.61 | 58.20    |

The surface morphologies of biochars are shown in Figure 3. The SEM image of BB300 revealed rod bundles with plicated surface and irregular small pieces, and BB400 also showed similar morphology. The SEM image of BB500 indicated that the rod bundles were partly destructed and became incompact. While the pyrolysis temperature was further raised to 600 °C, the rod bundles were almost destructed to small pieces and irregular sheets with some small strips.

**3.2. Effects of pH and Pyrolysis Temperature**

The adsorption of Cr(VI) can be significantly influenced by pH, due to its effects on (1) distribution of surface charge and ionic state of functional groups on the biochar surface [23,44,45], (2) formation species and oxidizing ability of chromates [22,44,46]. As seen in Figure 4, the Cr(VI) adsorption efficiencies of all banana pseudostem biochars generally increased with decreasing initial pH of Cr(VI) solution, and the maximum adsorption capacity was obtained at pH 1, which fully indicated that the biochars were more active under acidic conditions for Cr(VI) adsorption. For example, the adsorption of BB300 increased gradually from 6.94 mg/g to 18.36 mg/g with the decrease of pH from 5.0 to 3.0, then increased sharply to 189.89 mg/g as the pH decreased to 1.0. This was caused by the formation of massive positive charges on biochar surface and $\text{HCrO}_2^-$ in the solution at lower pH, which was favourable for the electron interaction between the biochars and chromate anions. Moreover, the oxidizing ability of chromate was significantly improved at low pH, leading to remove more Cr(VI) by reduction [22,23,42].

As a critical parameter influencing the physicochemical properties of biochars, pyrolysis temperature was found to play an important role in Cr(VI) adsorption capabilities of banana
pseudostem biochars. The adsorption capability of the biochars remarkably increased with increase of pyrolysis temperature from 200 to 300 °C, then decreased with further increase of pyrolysis temperature. The best adsorption capacity of BB300 was up to 189.89 mg/g at pH 1 and 25 °C (Figure 4). There were not significant changes in the surface area and microporosity of biochars altering pyrolysis temperature (Table 1). Whereas the content of O-containing groups which can be efficient factors for improving the redox, ion exchange and electron interaction abilities of biochars obviously reduced with the increase of pyrolysis temperature (Table 2). Thus, the decrease of adsorption capability with increase of temperature from 300 to 600 °C might be resulted from loss of surface O-containing groups. Similar results were also observed by Zhou et al. and Chen et al. [23,42]. As the adsorption capabilities of BB200 and BB600 were almost the same, the biochars prepared at different temperatures ranged from 300 to 600 °C were selected to study adsorption kinetics and isotherms.

![Figure 4](image-url)  
**Figure 4.** The effect of pH and pyrolysis temperature on adsorption capability (initial Cr(VI) concentration = 200 mg/L; adsorbent dose = 1 g/L; temperature = 25 °C, contact time = 24 h).

### 3.3. Adsorption Kinetics

The adsorption kinetics were studied at pH 2 and described by the pseudo-first-order (Equation (2)) and pseudo-second-order (Equation (3)) models [47]:

\[
\ln(q_e - q_t) = \ln q_e - k_1 t \tag{2}
\]

\[
\frac{t}{q_t} = \frac{1}{k_2 q_e^2} + \frac{1}{q_e} \tag{3}
\]

where \(q_e\) and \(q_t\) (mg/g) are the equilibrium and instant adsorption capacities, respectively. \(k_1\) (h\(^{-1}\)) and \(k_2\) (g mg\(^{-1}\) h\(^{-1}\)) are the corresponding adsorption rate constants.

The adsorption kinetics of banana pseudostem biochars are presented in Figure 5. The adsorption of Cr(VI) gently increased with contact time and almost reached equilibrium after 48 h. The calculated results of pseudo-first-order and pseudo-second-order rate equations are illustrated in Table 3. The correlation coefficients (R\(^2\)) of the pseudo-second-order kinetic model were generally better than that of pseudo-first-order kinetic model. Meanwhile, the \(q_e\) values calculated from pseudo-second-order equation were more consistent with the experimental results. These supported that the adsorption of Cr(VI) by banana pseudostem biochars was a chemiadsorption process. Adsorption kinetic studies of BB300 at different temperatures showed that Cr(VI) adsorption efficiencies of the biochars increased greatly with temperature (Figure 6), suggesting the process was endothermic and high temperature was favorable for chromium removal, which might be due to more facile reduction of Cr(VI) by biochars at higher temperature [48,49]. The results also revealed that the adsorption kinetics were better described by pseudo-second-order model than pseudo-first-order model (Table 4). Additionally, comparing with the experimental results, the higher calculated \(q_e\) results by pseudo-second-order model seemed more reasonable, as the adsorptions might not reach equilibrium within 48 h.
Table 3. The fitting results of Cr(VI) adsorption using pseudo-first-order and pseudo-second-order models.

|                    | BB300 | BB400 | BB500 | BB600 |
|--------------------|-------|-------|-------|-------|
| $q_{e,exp}$        | 102.45| 72.21 | 42.01 | 20.58 |
| $q_{e,cal}$        | 93.03 | 50.25 | 32.70 | 11.35 |
| $k_1$ (×10^{-2})   | 7.72  | 8.51  | 6.74  | 5.63  |
| $R^2$              | 0.96073 | 0.98212 | 0.90456 | 0.86206 |

Table 4. The fitting results of Cr(VI) adsorption on BB300 at different temperatures using pseudo-first-order and pseudo-second-order models.

|                    | 15 °C | 25 °C | 35 °C | 45 °C |
|--------------------|-------|-------|-------|-------|
| $q_{e,exp}$        | 72.04 | 102.45| 116.50| 156.56|
| $q_{e,cal}$        | 67.63 | 94.36 | 128.17| 135.94|
| $k_2$ (×10^{-3})   | 9.04  | 7.78  | 1.20  | 6.90  |
| $R^2$              | 0.96007 | 0.95519 | 0.96748 | 0.97669 |

![Figure 5](image1.jpg)  
Figure 5. The adsorption kinetics of biochar for Cr(VI). AC is the adsorption curve of Cr(VI); PFOM is the pseudo-first-order model; PSOM is the pseudo-second-order model (initial Cr(VI) concentration = 200 mg/L; adsorbent dose = 1 g/L; pH = 2; temperature = 25 °C).

![Figure 6](image2.jpg)  
Figure 6. The adsorption kinetics of BB300 for Cr(VI) at different temperatures. AC is the adsorption curve of Cr(VI); PFOM is the pseudo-first-order model; PSOM is the pseudo-second-order model (initial Cr(VI) concentration = 200 mg/L; adsorbent dose = 1 g/L; pH = 2).
3.4. Adsorption Isotherms

The Langmuir [50] and Freundlich models [51] were used to fit the isothermal adsorption data. These two models were described as Equations (4) and (5) respectively:

\[
q_e = \frac{K_L q_{\text{max}} C_e}{1 + K_L C_e} \quad (R_L = \frac{1}{1 + K_L C_0}) \tag{4}
\]

\[
q_e = K_F C_e^n \tag{5}
\]

where \(C_e\) is the equilibrium concentrations of Cr(VI) in solutions (mg/L); \(q_e\) and \(q_{\text{max}}\) (mg/g) are the equilibrium and maximum adsorption capacities of Cr(VI), respectively; The \(R_L\) parameter is the dimensionless adsorption factor; \(C_0\) (mg/L) is the initial concentration of Cr(VI) in the solution; \(K_L\) (L/mg) is the Langmuir constant related to binding energy. \(K_F\) ((mg/g) (mg/L)^{-n}) is the Freundlich constant related to adsorption capacity and \(n\) is an indicator of the adsorption intensity.

Batch adsorption isotherms are shown as Figure 7. The equilibrium adsorption capacities of all biochars generally increased with concentration of Cr(VI) and reached maximum, the maximum adsorption capacities of BB300, BB400, BB500 and BB600 were 125.44 mg/g, 80.33 mg/g, 43.47 mg/g and 21.53 mg/g respectively. The adsorption isotherms was better fitted by Langmuir model than Freundlich model (Table 5), indicating monolayer adsorption onto homogeneous surface with limited active sites was a dominating process for Cr(VI) removal by banana pseudostem biochars [52,53]. The average value of adsorption factor (\(R_L\)) of Langmuir isotherm was less than 1, suggesting that the removal of Cr(VI) by the biochars was a highly efficient process.

![Figure 7](image-url)  
*Figure 7.* The adsorption isotherms of biochars for Cr(VI). (a) is the Langmuir model; (b) is the Freundlich model (initial Cr(VI) concentration = 25–600 mg/L; adsorbent dose = 1 g/L; pH = 2; contact time = 48 h).

|       | BB300 | BB400 | BB500 | BB600 |
|-------|-------|-------|-------|-------|
| \(q_{\text{max}}\) (mg/g) | 125.44 | 80.33 | 43.47 | 21.53 |
| \(K_L\) (L/mg) | 0.1024 | 0.1910 | 0.3929 | 0.0796 |
| \(R^2\)     | 0.9977 | 0.9996 | 0.9999 | 0.9997 |

|       | BB300 | BB400 | BB500 | BB600 |
|-------|-------|-------|-------|-------|
| \(K_F\) (mg/g) (mg/L)^{-n} | 42.64 | 39.07 | 30.25 | 10.67 |
| \(n\) | 0.185 | 0.124 | 0.062 | 0.115 |
| \(R^2\) | 0.9813 | 0.9609 | 0.8030 | 0.9554 |

3.5. Adsorption Mechanism

The XPS and FTIR spectrometers were also performed after adsorption for exploring the Cr(VI) adsorption mechanism, as shown in Figures 8 and 9 respectively. The XPS results of BB300 after
adsorption revealed that Cr(III) was primarily immobilized on the surface of BB300 with a small amount of Cr(VI) (Cr(III)/Cr(VI) = 97.51/2.49). The content of O-containing groups on the surface slightly increased and both C–O and COO obviously increased (Table 2). In contrast, the content of C=O and CO$_2$– sharply reduced. These results indicated that reduction of Cr(VI) to Cr(III) [54–56] and adsorption of both Cr(III) and Cr(VI) onto the biochar surface were involved during the process. Additionally, the FTIR showed that the COO peak at 1686 cm$^{-1}$, aromatic C=C peak at 1429 cm$^{-1}$ and C–O peak at 1109 cm$^{-1}$ obviously increased for BB300 and BB400 after adsorption, implying that the biochars were oxidized by Cr(VI) and COO, C–O and aromatic ring were formed. Hence, the reduction of Cr(VI) by BB300 and BB400 during the adsorption was caused by oxidation and aromatization of the biochars. Differently, the aromatic C=C peak at 1421 cm$^{-1}$ diminished and C=O peak at 1620 cm$^{-1}$ increased for BB500 and BB600 after adsorption, indicating the reduction of Cr(VI) by BB500 and BB600 resulted from oxidation and dearomatization of the biochars. Moreover, adsorption kinetics were described well by pseudo-second-order model, suggesting the removal of Cr(VI) by the biochars was a chemical process. As a result, the decrease of adsorption capability of banana pseudostem biochar with increasing pyrolysis temperature was explained by the lost of surface O-containing groups which played an important role in reduction of Cr(VI). In general, the adsorption of Cr(VI) was mainly attributed to reduction of Cr(VI) to Cr(III) followed by complexation and ion exchange rather than electron interaction with biochar, as the biochar surface was probably positive at pH 2, hindering the electron interaction of Cr(III) cation with biochar. The precipitation of Cr(OH)$_3$ on the surface of biochar also could be ruled out, as the predominant species of trivalent chromium was Cr(III) cation at low pH [57].

![Figure 8. The XPS C 1s and Cr 2p scans and peaks fitting for BB300 after adsorption.](image)

![Figure 9. The FTIR spectra of biochars before and after adsorption of Cr(VI). Spectra (a) were before adsorption; Spectra (b) were after adsorption.](image)
4. Conclusions

The Cr(VI) adsorption capabilities of banana pseudostem biochar were highly affected by pyrolysis temperature. Low pH and high adsorption temperature were favourable for adsorption of Cr(VI). The biochar prepared at 300 °C exhibited best adsorption capability with 125.44 mg/g maximum adsorption capacity at pH 2 and 25 °C. The high adsorption capability of the biochar might be attributed to high content of oxygen-containing functional groups on surface. The Cr(VI) adsorption by BB300 and BB400 resulted from their oxidation and aromatization with ion exchange and complexation, whereas the Cr(VI) adsorption by BB500 and BB600 was ascribed to their oxidation and dearomatization followed by ion exchange and complexation. As the feedstock is large-scale available and highly renewable biowaste, banana pseudostem biochar is a low-cost adsorbent and expected to have a practical application for remediation of chromium-polluted water in the future.

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