Restoration of Bare Incrusted Soils in the Sahel Region of Burkina Faso

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Abstract: Bare incrusted soils are a degradation stage often encountered in the Sahel zone. Our study documents the success of restoration (= regreening) experiments using deep ploughing in an experimental site south of Gorom-Gorom in the Oudalan province of Burkina Faso. We used phytosociological relevés and maximum likelihood classifications of digital photography to analyze changes in vegetation. Plant cover in treated plots was found to be about 20 times higher than in control plots, mean species richness more than twice as high. Therefore, this promising restoration method should be tested also in other Sahelian regions. Our approach to combine phytosociological relevés and maximum likelihood classifications of digital photography proved to be very useful.

Key words: biodiversity, rangeland regeneration, soil crusts, Vallerani system

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1 Introduction

The trans-African belt of the Sahel is located at the southern fringe of the Sahara, the world’s largest desert. The vegetation in its North is characterized by annuals and herbaceous perennials, while the South is dominated by Acacia savannas (White 1983). Although some agriculture is possible in areas with a comparably good water availability (e.g. millet on dunes), the main land use is animal husbandry, usually pastoralism with cattle and small ruminants (Fig. 1). In the Oudalan province and the entire Sahel region of Burkina Faso, grazing pressure and resulting erosion has led to the formation of bare incrusted soils called zipelé in Burkina Faso or hardé in Chad (Zougmore et al. 2003).

These soils have a reduced permeability for water which prevents their regreening. To break out of this vicious circle, different approaches have been developed, including termite- and mulch-mediated rehabilitation (Mando et al. 1999), the Zaï system (Roose et al. 1999; Ghana et al. 2006; Sadowsko et al. 2008) and the Vallerani system (Pari & Antinori 2001; Malagnoux 2009). Here we present a case study documenting the restoration success resulting from the application of the latter system in the Sahelian region of Burkina Faso. The aim of the experiment was to regreen the bare crusts, i.e. increase plant cover in as short time as possible by simple and locally available means.

2 Methods

The study area is situated in the province of Oudalan (Burkina Faso), south of Gorom-Gorom, near the road to Dori. There the rainy season lasts three to four months. In the last decade, average annual rainfall in this area was found to be between 400 and 500 mm (Wittig et al. 2006), showing great local and interannual variation. In the year of investigation, an annual precipitation amount of c. 600 mm was observed (supplementary material to Nielsen & Reenberg 2010).

The incrusted soils are found in the glacis area. The natural vegetation of the glacis is formed by thornbush savannas and tibergush, a mosaic of dense shrubby vegetation and bare soil, appearing on slight slopes and certain soil conditions. Today in Burkina Faso, tibergush only remains in re-
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Fig. 1: Generalised profile of the Sahelian landscape in northern Burkina Faso (from Wittig et al. 2002; amplified and translated into English); vegetation: selected dominant units (for more details see Müller 2003, Ouédraogo et al. 2005)

Fig. 1: Profil généralisé du paysage sahélien au nord du Burkina Faso; végétation: unités dominantes sélectionnées.

mote areas. In the vicinity of settlements it has totally disappeared due to overexploitation. In these areas, not incrusted soils are covered by the Schoenefeldia gracilis community with a coverage of 70 to 90 % (Müller 2003). On incrusted soils however, vegetation cover is generally below 5 %.

Due to locally varying transport and accumulation processes, the soils of the glacis are very inhomogeneous. Soil incrustation is caused by sheet wash which leads to the formation of run-off crusts. Incrusted soils show only very little infiltration (Albert et al. 2004). Moreover, the small part of the precipitation that is infiltrated is not available for plants but strongly retained in the micropores of the argillaceous material (Claude et al. 1991). As long as an intact tiger bush exists, runoff is retained by the bush stripes which usually grow perpendicular to the sheet flow direction.

Fig. 2: Photographic records of restoration treatment (right) and control plot (left).
Fig. 2: Documentation photographique des placeaux de restauration (droite) et contrôles (gauche).
At the beginning of the rainy season of the year 2002, a degraded glacis was partly deep-ploughed by ADRA Gorom-Gorom (Adventist Development and Relief Agency) using the particular ploughs that are characteristic of the Vallerani system. The remaining part was left undisturbed for control. Although animals were supposed to be excluded from the area, we observed some goats drinking from the puddles of the plough traces. We randomly placed ten rectangular plots (size 5.0 m x 3.5 m), both in the ploughed and the control area. In the deep ploughed area the longer side of the rectangles ran along the direction of the plough lines, the shorter side perpendicular to these, with the plough line exactly in its centre. We recorded plant cover for each species in percent, since the often used cover classes of Braun-Blanquet (1964) or of Londo (1976) would not have been suitable for the description of the very sparse vegetation and consequently would have corrupted our results.

Within the area of the botanical relevé we randomly placed five digital photos at a height of 1 m from the ground and a standard focal length (Fig. 2). A maximum likelihood algorithm was used to classify the pixels into vegetation and bare soil (Fig. 3). We tested the classification with visual verification of 1000 random pixels and had an error rate of only 0.3%.

Relevé data was stored in VegDa (Schmidt 2006). Similarity in species composition among relevés was analyzed with cluster analysis using PCORD 5.0. Sorensen coefficient (Sørensen 1948) was used as similarity measure and clusters were defined by the flexible beta grouping method with β = -0.25. Diagnostic species were identified using the fidelity measure \( q \) (fidelity∗100) that compares the frequency of a species within a group with the constancy of the same species in all other groups within the dataset. All diagnostic species had \( q \)-value higher than 30 and significance \( p<0.05 \) according to Fisher’s Exact Test.

Differences in the number of species and evenness, as measure for equitability (Pielou 1969); between ploughed and unploughed sites were compared with Univariate Analysis of Variance (ANOVA) using the software SPSS 15 (SPSS Inc., Chicago IL, USA).

### Results

Vegetation cover and species richness were significantly higher \( (p < 0.001) \) in deep plough treated plots than in the control plots: the mean vegetation cover was approximately 20 times higher (Fig. 4), mean species richness more than twice as high (Fig. 5). Also evenness is significantly higher in the ploughed plots (\( p < 0.001 \)).
The program JUICE (Tichy 2002) identifies six species as significantly differentiating the control plots from the ploughed (s. Table 1): Alysicarpus ovalifolius, Amaranthus graecizans, Curcumin melo, Eleusine indica, Mollugo nudicaulis, Sporobulus micropotus and Zornia glochidiata. The grass Aristida spec. was only found in the control plots.

Vegetation of control plots and deep ploughed plots can both be identified as Schoenefeldia gracilis-community. The photos (Fig. 2) illustrate how rainwater was harvested in the plough traces and herbaceous vegetation reappeared around these pools.

Table 1: Floristic composition of control plots and deep ploughed plots*  
Tableau 1: Composition floristique des placeaux labourés et contrôles

| Relevé/Plot No. | 12 | 20 | 13 | 19 | 14 | 15 | 16 | 17 | 11 | 18 | 4 | 5 | 3 | 1 | 2 | 9 | 10 | 8 | 6 | 7 |
|----------------|----|----|----|----|----|----|----|----|----|----|---|---|---|---|---|---|---|---|---|---|
| Plot type      | control | deep ploughed |
| Vegetation cover (%) | <1 | 1 | <1 | 1 | 1 | <1 | <1 | 10 | 5 | 15 | 10 | 10 | 8 | 5 | 20 | 15 | 10 |
| Number of species | 4 | 5 | 6 | 5 | 8 | 5 | 5 | 4 | 3 | 15 | 10 | 14 | 9 | 14 | 10 | 13 | 9 | 11 | 13 |

**Differential species of deep ploughed plots**
- Alysicarpus ovalifolius
- Zornia glochidiata
- Mollugo nudicaulis
- Amaranthus graecizans
- Cucumin melo
- Waltheria indica
- Sporobulus micropotus

**Differential species of control plots**
- Aristida spec.
- Schoenefeldia gracilis
- Corchorus tridens
- Indigofera senegalensis
- Spermacoce chaetocephala
- Cassia obtusifolia
- Panicum laetum
- Acacia tortilis var. raddiana (seedl.)
- Ledebouria sudanica
- Tribulus terestris
- Chenchus biflorus
- Digitaria ciliaris
- Tragus racemosus
- Boerhavia repens viscosa
- Euphorbia forsskallii
- Chloris prieurii
- Poaceae indet.
- Balanites aegyptiaca (seedl.)
- Eragrostis pilosa
- Eragrostis ciliatensis
- Leptadenia hastata (seedl.)
- Dactyloctenium aegyptium
- Cyperus iria
- Echinochloa colona

* Nomenclature follows African Plants Database (http://www.ville-ge.ch/musinfo/bd/cjb/africa/)*
4 Discussion

As not all soils of the glacis in the study area are incrusted, there are enough reference sites to estimate the vegetation which should be regarded as restoration target. The method described by Lane & Texler (2009) for the absence of reference sites therefore was not needed. As Table 1 and 2 show, deep ploughing proved to be highly effective in restoring vegetation cover as well as species richness. Part of the species richness might be due to the diversity of microhabitats created by the ploughing, but natural Sahelian savanna vegetation of the Schoenefeldia gracilis-type has a comparable species richness (see Table 2). As demanded by Hobbs (2007), the goals were effective and realistic. Also the socio-economic aspects were considered: People in this area are not able to buy fertilizer and they cannot afford to reduce grazing intensity for a longer time in large areas. Compost, recommended by Curtis & Claassen (2009) as regeneration tool, also is generally not available in the Sahel. Restoration by grazing management which in many cases is recommendable (Papanastasis 2009) cannot work in this area because it will not lead to a reduction of the incrustation. Considering the implication of climate change, any restoration method has to be checked for its applicability in the changed biophysical conditions of the future (Harris et al. 2006). For deep ploughing one can be sure that this method will fit for the Sahelian region also under the conditions predicted by the IPPC.

Our approach to use digital photography appears to be a promising tool when it comes to quantification of vegetation cover, and maybe even biomass (see, e.g., Flombaum & Sala 2009). It is easy to use, time-efficient and precise (which is often not the case with estimation methods: see, e.g., Leps & Hadincova 1992; Klimes 2003), but is presently not applicable for studies where individual species have to be differentiated.

5 Implications for Practice

- Our study shows that Vallerani ploughs can be effectively used for restoration of vegetation on incrusted soils in semi-arid areas.
- Classification of digital photographs is an easy and accurate way to measure the success of regreening measures.

6 Outlook

Our promising results demand for a prolongation of the observation period and an extension of the experiment to other Sahelian areas. Considering the recent sahelisation of large parts of the Sudanian zone (Wittig et al. 2006, 2007), the method should also be tested in Subsahelian and North Sudanian areas.

The regreening achieved in the experiments is only a first step. For many areas of the Sahel, the final aim should be the regeneration of the water harvesting tiger bush. To reach this aim the exclusion of grazing for some years is a condition sine qua non.

**Table 2: Floristic composition (% constancy) of incrusted plots (1), deep ploughed (formerly incrusted) plots (2) and of not incrusted areas of the glacis (3)**

| Plot type                        | 1     | 2   | 3    |
|----------------------------------|-------|-----|------|
| Number of plots                  | 10    | 10  | 69   |
| Mean vegetation cover (%)        | <1    | 11  | 64   |
| Mean species number              | 5     | 11  | 13   |
| Schoenefeldia gracilis           | 70    | 100 | 96   |
| Panicum laetum                   | 30    | 70  | 59   |
| Spermacoce chaetocephala         | 50    | 60  | 68   |
| Corchorus tridens                | 10    | 100 | 39   |
| Cassia obtusifolia               | 40    | 60  | 14   |
| Acacia tortilis var. raddiana    | 20    | 60  | 14   |
| Digitaria ciliaris               | 10    | 40  | 33   |
| Mollugo nudicaulis               | 10    | 80  | 36   |
| Tribulus terrestris              | 40    | 10  | 46   |
| Boehavria repens viscosa        | 20    | 10  | 38   |
| Ledebouria sudanica              | 40    | 20  | 9    |
| Leptadenia hastata (seedling)    | 30    | 10  | 12   |
| Euphorbia forskalii              | 10    | 30  | 15   |
| Alysicarpus ovalifolius          | 80    | 75  |      |
| Zornia glochidiata               | 70    | 65  |      |
| Enteropogon prieurii             | 30    | 62  |      |
| Cenchrus biflorus                | 20    | 54  |      |
| Indigofera senegalensis          | 90    | 17  |      |
| Waldheria indica                 | 40    | 4   |      |
| Sporobolus micropus              | 40    | 4   |      |
| Amaranthus graecizans            | 50    | 1   |      |
| Dactyloctenium aegyptium         | 10    | 84  |      |
| Eragrostis pilosa                | 10    | 32  |      |
| Aristida spec. / A. mutabilis    | 50    | 29  |      |
| Cucumis melo                     | 50    |     |      |
| Tragus berteronianus             | 30    |     |      |
| Indigofera aspera                | 30    |     |      |
| Aristida fusiculata              |      | 29  |      |

* only species with a constancy >20 %

sources: 1 and 2: this paper (Table 1); 3: Müller (2003; Table 17, No. 24-92)

When the tiger bush is restored it will not only serve as a water harvesting system (Valentin & d’Herbés 1999), but can be used for grazing and as source for wood. However, to avoid re-degradation, a sustainable management has to be implemented. Regeneration of tiger bush may be also of interest for carbon-sequestration, because woody biomass production has proved to equal that of the forest in more humid southern zones and to exceed that of industrial plantations in the same area (Leprun 1999).
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