Effectiveness of low impact development for urban inundation risk mitigation under different scenarios: a case study in Shenzhen, China

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Abstract. The increase in impervious surfaces associated with rapid urbanization is one of the main causes of urban inundation. Low impact development (LID) practices have been studied for mitigation of urban inundation. This study used a hydrodynamic inundation model, coupling SWMM (Storm Water Management Model) and IFMS Urban (Integrated Urban Flood Modelling System), to assess the effectiveness of LID under different scenarios and hazard levels. The results showed that LID practices can effectively reduce urban inundation. The maximum inundation depth was reduced by 3–29 %, average inundation areas were reduced by 7–55 %, and average inundation time was reduced by 0–43 % under the eight scenarios. The effectiveness of LID practices differed for the three hazard levels, with better mitigation of urban inundation at a low hazard level than at a high hazard level. Permeable pavement (PP) mitigated urban inundation better than green roof (GR) under the different scenarios and hazard levels. We found that more implementation area with LID was not necessarily more efficient and the scenario of 10 % PP + 10 % GR was more efficient for the study area than other scenarios. The results of this study can be used by local governments to provide suggestions for urban inundation control, disaster reduction and urban renewal.

1 Introduction

In recent years, urban stormwater inundation hazards have occurred frequently in major cities all over the world, leading to significant property damage in local areas (Bhattarai et al., 2016). In China, according to a report by the Ministry of Housing and Urban-Rural Development (MOHURD) in 2010, 62 % of 351 cities have suffered from inundation hazards, and 137 of these have had negative effects from urban floods on more than three occasions from 2008 to 2010. In 2012, 2013, 2014, and 2015, the number of cities that suffered urban inundation was 184, 234, 125, and 154, respectively, including Beijing, Shanghai, Guangzhou and Shenzhen. Urban inundation increasingly threatens the sustainable development of urban areas.
Rapid urbanization has become an important cause of frequent urban stormwater inundation, in addition to extreme precipitation and low standards for urban drainage infrastructure (Arnold, 1996; Beckers et al., 2013; Claessens et al., 2006; Zahmatkesh et al., 2015b). Rapid expansion of cities generally leads to an increase in impervious surfaces, which makes the hydrological characteristics of the urban surface change significantly (Arnold, 1996; Jacobson, 2011; Rose and Peters, 2001).

Impervious surfaces replace rivers, lakes, green spaces, and urban forests; weaken the flood control capability of the urban system; and change infiltration, evaporation, filtration, and storage (Hao et al., 2015; Jacobson, 2011; Meyer, 2001). The expansion of impervious areas accelerates rainwater convergence on urban surfaces, resulting in increased runoff and peak flows (Hatt et al., 2004; Leopold et al., 1995; Liu et al., 2015). The increase in runoff and peak flows taxes urban drainage facilities and exacerbate the risk of urban inundation.

To solve the problem of urban inundation, scholars in China have suggested the “Sponge City” initiative, which allows cities to act as sponges to filtrate, purify, evaporate, and store rainwater (Mao et al., 2017; Sang and Yang, 2016). Low impact development (LID), an important development concept for sponge cities, has been applied in Sponge City construction (Luan et al., 2017); it is widely applied to reduce the impacts of urban inundation associated with rapid urbanization (Dietz and Clausen, 2008; Dietz, 2007; Xia et al., 2017; Zahmatkesh et al., 2015a). LID is a stormwater management strategy that uses microscale and localized practices to control the runoff and pollution caused by a storm (Damodaram et al., 2010; EPA, 2000; HUD, 2003). Since the 1990s, LID practices have been widely used in countries in Europe, the United States of America, and other developed countries. LID practices include PP, GR, bioretention, swales, infiltration wells/trenches, infiltrating wetlands, and rain barrels (Hunt et al., 2010).

The hydrological effectiveness of LID practices has been researched through field and laboratory studies (Abbot and Comino-Mateos, 2003; Berndtsson, 2010; Davis, 2008; Davis et al., 2012; Fassman and Blackbourn, 2010). For example, Hood et al. (2007) monitored low impact residential development and traditional residential development in Waterford, Connecticut, USA, and found that LID practices helped lower runoff, peak flows, and discharge volumes. Dreelin et al. (2006) designed a test to compare the performance of asphalt and PP parking lots in Athens, Georgia, USA, and their results showed that the porous parking lot contributed 93% less runoff than the asphalt lot during natural storm events. Bliss et al. (2009) constructed and monitored a GR in Pittsburgh, Pennsylvania, USA and reported that the GR reduced runoff by up to 70% and reduced peak flows by 5–70%; the hydrograph was delayed by several hours more than a normal roof for the same building.

Many scholars have focused on simulations at a large scale, such as watersheds (Ahiablame et al., 2012; Dietz and Clausen, 2008; Roy et al., 2008; Salvadore et al., 2015), to explore the effectiveness of LID practices. For example, Palla and Gnecco (2015) reported that the LID combination of GR and PP decreased runoff and peak flows by 23% and 45%, respectively, and delayed the hydrograph by up to 19% at the urban catchment scale. Trinh and Chui (2013) conducted a simulation and found that GR could reduce the peak flows by 50% and delay the hydrograph by 2 hours, bio-retention (BR) systems could reduce the peak flows by 50%, and the combined GR and BR systems could reduce the peak flows to a pre-urbanized level. Morsy et al. (2016) reported that rain gardens can mitigate runoff by approximately 15, 27, and 38% for 2-, 5-, and 10-year storm events, respectively, which reduced the watersheds flood risk. Ahiablame et al. (2013) assessed the effectiveness of rain
barrels/cisterns and PP in two urbanized watersheds near Indianapolis, Indiana, USA; by using simulations, they found that
LID practices reduced runoff and pollutant loads; they listed some LID combinations that are good retrofitting options for local
areas.
Peak flows reduction, runoff reduction, and hydrograph delays are widely used indexes for evaluating the performance of LID
practices (Ahiablame and Shakya, 2016; Qin et al., 2013; Zhang et al., 2016). However, these indexes are not intuitive, and
the performance of LID practices for urban inundation is more useful for local residents, such as providing a guide for their
travel behaviour. Some 1D-2D models have been applied for flood management, such as ESTRY-TUFLOW (Fewtrell et al.,
2011), InfoWorks ICM (Russo et al., 2015) and MIKE FLOOD (Loewe et al., 2017). However, most of these models have a
cost, which limits their application. In recent years, as an open source and free model, SWMM has been coupled with other
models, such as BreZo (Burns et al., 2015) and LISFLOOD-FP (Wu et al., 2017), to simulate urban inundation, which means
that the coupled models based on SWMM is needed in future research.
The goal of this study was to evaluate the effectiveness of LID practices to mitigate urban inundation in urban watershed using
a case study. The specific objectives were to establish a 1D-2D hydrodynamic model that coupled SWMM and IFMS Urban,
evaluate the effectiveness of LID practices under different scenarios and hazard levels, and explore the efficiency of the LID
scenarios. We intended this study to enrich LID inundation mitigation research at the urban watershed scale and to provide a
reference for urban stormwater management and inundation mitigation for local governments.

2 Materials and methodology

2.1 Study site
Shenzhen is in the coastal area of Guangdong Province in southern China (Figure 1). It has a subtropical maritime monsoon
climate; Shenzhen is hot and rainy in summer and mild in winter, and the average annual rainfall is 1837 mm. April to
September is the rainy season in Shenzhen, and during this period, precipitation is concentrated and stormwater overflows are
frequent. There were 38 rainstorm days (95 % of the year) in 2017 and the average rainfall was 170–350 mm every month
during this period. Accordingly, urban inundation was particularly serious in this period; it caused loss of life and economic
losses for local residents.
The study site was located in Guangming New District of Shenzhen, China, and it is in the Maozhou River Basin (Figure 1).
The total area of our study site was 37.68 km², of which 69.8 % was impervious surfaces. Guangming New District was
selected as the first pilot area for LID practices in Shenzhen in October 2011 because of the intensity of its inundation disasters.
There is a need to research the effectiveness of LID on urban inundation mitigation in this area.

2.2 Data
The model input data included inundation, land use, a digital elevation model (DEM), weather, and pipe network data. The
land use data (2013) and pipe network data were provided by the Shenzhen government. We generalized the original data and
divided the study area into water, low density construction land, high density construction land, bare land, woodland, grassland, and agricultural land using remote sensing images (Figure 1). The DEM of the study area (Figure 2) was downloaded from the Geospatial Data Cloud (30 m resolution). The weather data were sourced from the Shenzhen Meteorological Data System (https://data.szmb.gov.cn/). According to the integrity and availability of data, we chose two representative heavy rainstorm event datasets, 11 May 2014 and 10 May 2016 (Figure 3) for model simulation, which included the complete volume of rainfall every hour. The corresponding inundation data were obtained from the Shenzhen SanFang (flood, drought, and wind defence) headquarters and the Guangming New District Urban Construction Bureau. We simplified the drainage data for building the model because the urban pipe network is intricate and substantial: add nodes when the pipeline is too long; keep or add the nodes that change the diameter and slope of pipeline; keep the parallel pipelines and nodes on both sides of the roads; and delete the nodes and pipelines that independent of this model. Finally, the 4502 pipelines and 1175 nodes in this study were generalized to 597 pipelines and 653 nodes, including 56 outlets and 597 inspection nodes (Figure 2).

### 2.3 SWMM and IFMS Urban models

Developed by the United States Environmental Protection Agency (US EPA), SWMM is an open-source model that can simulate dynamic runoff quantity and quality from urban areas, and it has been widely used to simulate the hydrologic performance of LID practices (Rossman, 2010; Wu et al., 2013). However, SWMM cannot simulate the spatial and temporal distributions of surface inundation. Recently, some scholars have conducted simulations using secondary developments of this software (Seyoum et al., 2012; Son et al., 2016; Zhu et al., 2016). We expected that this application would be difficult to use in our study area due to differences in computer programming. Coupling a model with SWMM for 2D simulation is another way to simulate the spatial distribution of urban inundation (Huong and Pathirana, 2013; Wu et al., 2017).

The Integrated Urban Flood Modeling System (IFMS Urban) was developed by the China Institute of Water Resources and Hydropower Research (IWHR) in cooperation with other institutions. Based on the simulated results from SWMM, IFMS Urban can simulate the temporal and spatial distribution of urban inundation, and it is compatible with ArcGIS and SWMM. Data conversion and model coupling are accomplished in IFMS Urban, and it does not need additional software programming, which is convenient for researchers and non-expert users.

Therefore, we coupled SWMM and IFMS Urban in this study to simulate urban inundation. SWMM was applied to construct a 1D sewer model. The study area was simplified to 577 sub-catchments, 597 pipelines, and 653 nodes. Details of model building and of SWMM’s parameters can be found in many published studies (e.g., Rossman, 2010; Qin et al., 2013; Wu et al., 2017). Model coupling occurred in IFMS Urban. First, an unstructured 2D grid model was meshed with an average cell size of 15 m; second, ground elevations were assigned to each grid; finally, each node was linked with a corresponding grid for water exchange, and the distribution of surface inundation was calculated with 2D shallow water equations. The coupled model had the advantages of SWMM and IFMS Urban, and could be applied to simulate urban inundation and evaluate the performance of LID practices.
2.4 Scenarios of LID combinations for simulation

Considering the feasibility and representativeness of LID practices for urban inundation mitigation, we chose two types of LID practices, GR and PP, to simulate and explore their effectiveness for mitigation of urban inundation. The parameters for PP and GR are listed in Table 1, which were designed based on SWMM requirements and LID research (Ahiablame and Shakya, 2016; Chui et al., 2016; Kong et al., 2017; Qin et al., 2013). Through remote sensing images and field investigations, we found that urban villages have diverse roof structures and shapes, which makes it difficult to implement green roofs. Therefore, we established principles for the implementation of LID practices: GR can only be built on low density construction land, and PP can be built on low and high construction land and on some streets. According to these principles, the available area for PP and GR was 5.95 km² and 8.92 km², respectively. We set a series of proportions from 5% to 100% for the density of different types of LID combinations, and a benchmark and eight scenarios are designed below:

Benchmark: No LID practices
Scenario 1 (S1): 5% GR + 5% PP
Scenario 2 (S2): 10% GR + 10% PP
Scenario 3 (S3): 25% GR + 25% PP
Scenario 4 (S4): 50% GR + 50% PP
Scenario 5 (S5): 75% GR + 75% PP
Scenario 6 (S6): 100% GR + 100% PP
Scenario 7 (S7): 100% PP
Scenario 8 (S8): 100% GR

3 Results

3.1 Model calibration and validation

The coupled model was calibrated using rainfall and inundation data from 11 May 2014. Based on the relevant literature and the SWMM manual, we determined the final SWMM parameters (Table 2) through several calibration iterations. From the final calibration results (Table 3), we found that, except for inundation site Gm 20, the absolute value of the maximum inundation depth between the observed and simulated value was approximately 0–0.14 m and the relative error was ranged from 0–30%.

To further confirm the applicability of the model, the rainfall and inundation data on 10 May 2016 was chosen to validate the coupled model. Three valid datasets were simulated with the coupled model using observed urban inundation data on 10 May 2016 from the Guangming New District Urban Construction Bureau. The results showed that the absolute values of the differences between the observed and simulated maximum inundation depths were 0.04 m (Gm 11), 0.05 m (Gm 12) and 0.02 m (Gm 20), and the relative errors were 20, 7, and 5%, respectively. In this study, the relative error of calibration were a little
higher, while the relative errors of validation were 5–20 %, which met the requirements of the Standard for Hydrologic Information and Hydrologic Forecasting in China (GBT_22482-2008).

3.2 Inundation depth under different scenarios

Figure 4 and Table 4 show the simulation results of inundation depths under different scenarios. Compared to the benchmark, the reduction rates of maximum inundation depth were 3, 7, 16, 22, 26, and 29 % under S 1 to S 6, respectively. The results for the 100 % PP and 100 % GR scenarios showed that PP and GR had approximately the same performance at the maximum inundation depth and both scenarios reduced maximum inundation by 14 %.

To further explore the impacts of LID practices on inundation mitigation, we set three hazard levels for the depth of urban inundation: low (< 0.2 m), medium (0.2–0.4 m), and high (≥ 0.4 m), based on the literature (Su et al., 2016) and observed data for the study area. Compared to the benchmark, the ranges of average depth reduction rates were 15–80, 5–54, and 4–40 % at low, medium and high hazard levels, respectively, for S 1 to S 6 (Figure 5a). Under different hazard levels, the average depth reduction rates increased from S 1 to S 6. The average depth reduction rates at the low level were 11, 26, 38, 44, 43, and 40 % higher than the high level under S 1 to S 6, respectively. These results suggest that most inundated areas could not be eliminated at the high level.

Figure 5a shows that the average depth reduction rates of 100 % PP and 100 % GR scenarios were between the 25 % GR + 25 % PP and 50 % GR + 50 % PP scenarios under different hazard levels. These results suggest that LID combinations may be more effective in reducing urban inundation than a single type of LID practice. Based on the comparison of the two LID practices, we found that the average depth reduction rates of the 100 % PP scenario were 67, 38 and 23 % at the low, medium and high levels, respectively. These were 6, 7, and 2 % higher than the average depth reduction rates of the 100 % GR scenario. These results suggest that PP may perform better than GR for reducing the depth of inundation.

3.3 Inundation areas under different scenarios

Figure 5b shows changes in the inundation area under different scenarios and hazard levels. Compared to the benchmark, the ranges of average area reduction rates were 6–53, 17–75, and 24–90 % at low, medium, and high levels, respectively, for S 1 to S 6. The inundation areas reduced at different hazard levels after the implementation of LID practices. The average area reduction rates at the high level were up to 24–90 %, which were greater than those at the low level. This likely occurred because, after the implementation of LID practices, the depth of inundation decreased and most inundated areas were downgraded from a high level to a medium level or a low level.

For the 100 % PP and 100 % GR scenarios, the reduction in the inundation areas was similar to the 25 % PP + 25 % GR scenario, which also suggested that LID combinations are more effective than single LID practice. The average area reduction rates for the 100 % PP scenario were 37, 65 and 67 % at the low, medium and high levels, respectively, which were 5, 9, and 0 % higher than those for the 100 % GR scenario.
3.4 Inundation time under different scenarios

Inundation time is another way to represent inundation risk. Table 5 shows that the inundation time for medium and high levels was longer than the inundation time for the low level under the same scenario, which reflects increased risk of inundation at medium and high levels. As the implementation area of LID increased, the average inundation time decreased under the three hazard levels. The 100 % PP and 100 % GR scenarios had lower inundation time than the 25 % PP + 25 % GR scenario, and the inundation time for the 100 % PP scenario was 1.3 h less than the inundation time for the 100 % GR scenario. Compared to the benchmark, the average inundation time under S 1 to S 3 changed slightly, while it decreased from S 4 to S 6. This result did not indicate that LID practices cannot decrease inundation time or that the model had errors. The inundation time decreased for all hazard levels, but for the low and medium levels, some areas inundated for a short-time were no longer flooded, which resulted in a different urban inundation area after the implementation of LID practices. Therefore, the average inundation time was longer than before LID practices were implemented at the low and medium levels. As LID practices were implemented, the average inundation time decreased continuously from 4.1 to 2.3 h under S 3 to S 6.

4 Discussion

4.1 Performance of PP and GR

Researching the effectiveness of LID practices for urban inundation mitigation is important for stormwater management. Our analysis showed that, although the implementation area of PP was less than GR, PP provided better urban inundation mitigation than GR. This result may have been due to differences in the LID parameters, but it may also have been caused by the PP’s more diffuse spatial pattern. To better identify the effects of parameters, we did a sensitivity analysis carried out by assuming a 50% increase in some parameters under S 7 and S 8, and the results showed that the inundation depth has great sensitivities to some parameters (Table 6). Under the permeable pavement scenario, the inundation decreases 15 %, 16 % and 18 % with thickness of pavement layer, thickness and void ratio of storage layer, respectively. Under the rain roof scenario, the inundation decreases 17 % and 19 % with thickness and porosity of soil layer, respectively. The results indicate that LID parameters might influence the effectiveness on inundation mitigation.

Indeed, except the LID parameters, there are some other factors, such as implementation area, spatial pattern, rainfall intensity and rainfall frequency that will influence the effectiveness, and these are the reasons why PP cannot always perform better and showed varying effectiveness in different studies (Ahiablame and Shakya, 2016; Hu et al., 2017; Qin et al., 2013; Zhang et al., 2016). However, under certain scenarios of this study, PP may be a good choice for local governments because of its effectiveness for stormwater management and its potential use for reconstruction in built-up areas. PP could be gradually applied to roads and parking lots, while GR is harder to implement in densely urbanized areas, especially in the urban villages.
4.2 Effectiveness at different hazard levels

At the high level, the average depth reduction rates decreased from 4% to 40%, and the average area reduction rates decreased from 24% to 90% under S1 to S6. These results showed that the inundation hazard eased at a high level with the implementation of LID practices. However, at the high level, the average depth reduction rates were still 11–40% lower and the average inundation time was 2.5–7 h longer when compared to the low level; this indicates that LID practices are more effective for urban inundation mitigation at a low hazard level. The hazard level analysis showed that although LID practices can downgrade the inundation hazard level to medium or low, most inundated areas cannot be eliminated at a high hazard level. This means that the inundation problem could not be resolved only with LID practices; and we should recognize the insufficiencies of LID practices.

4.3 Cost-effectiveness of LID practices

Under S1 to S6, the effectiveness of LID practices for urban inundation mitigation increased with more area implementing LID practices. However, Table 4 and Figure 5 showed that the reduction rates grew slowly with the increase of LID practices from 10% to 100%, which suggests that the efficiency of LID practices decreased from S2 to S6. A cost-effectiveness indicator (CEI) was used to better describe this phenomenon (Wu et al., 2017):

\[ \text{CEI} = \frac{R}{P}, \quad (1) \]

where R is the reduction rate of inundation depth and inundation area, and P is the proportion of LID practices which means the cost. Table 7 showed that the CEI increased from S1 to S2 and decreased from S2 to S6, which means that the efficiency of the 10% PP + 10% GR scenario was higher than other scenarios. And we can clearly find that the reduction rates of maximum inundation depth are 7, 16, 22, 26 and 29% from S2 to S6 and the CEI has reduced continuously, especially from S4 to S6. This indicates that wider implementation of LID practices may not lead to higher efficiency.

One of the causes behind the phenomenon is that LID practices can not control all the runoff of the watershed. Indeed, the runoff might not only come from sub-catchments around the inundation areas, but also come from other sub-catchments through the roads and pipe networks. And in this study, there are still some areas that can not implement LID practices. Therefore, the runoff from these areas can not be controlled by LID practices and directly influenced the effectiveness of inundation mitigation.

The phenomenon is common. In urban watershed, we could not transform all the roofs and roads to LID practices, and there are still some impervious covers that could influence the inundation that LID practices can not control. Therefore, we should recognize the insufficiencies of LID practices, and consider combine other measures such as restoring river systems, establishing urban wetlands, and improving urban drainage infrastructure to further promote the effectiveness on inundation mitigation.

Besides, properly implementing construction intensity of LID practices to achieve optimal efficiency in urban watershed will be very important for the construction of Sponge City.
4.4 Limitations and future studies

Lacking accurate data is a common limitation for most studies. In this study, highly accurate elevation data for the study area is confidential and difficult to obtain; therefore, the ground elevation of streets were interpolated from the dense nodes of the pipe network. This method may have affected the simulation results. Moreover, the accuracy of the coupled model could be further increased with more accurate observed data and information of infrastructure, such as drainage pump station and river channel. Another limitation was that the definition of the thresholds for hazard levels was not considered sufficiently in this study. The results for the three hazard levels would be different if the thresholds changed. Therefore, research on criteria and sensitivity analysis of thresholds is needed in the future. The influences of rainfall intensity and frequency were not considered in this study, which is related to the effectiveness of LID.

In China, urban inundation appears to be increasing, and LID practices could be efficient strategies for urban inundation mitigation. At present, most research has focussed on the area with LID practices and the effects on urban inundation mitigation. However, the spatial distribution and landscape patterns of LID practices also contribute to urban flooding mitigation (Giacomoni and Joseph, 2017; Kim and Park, 2016), but few studies have considered these variables. In addition, more studies should consider effective integration of LID practices into urban development (Chui et al., 2016), especially for places vulnerable to urban flooding.

5 Conclusion

This study constructed a 2D inundation model that coupled SWMM and IFMS Urban at the urban watershed scale; the model was used to evaluate the effectiveness of LID practices for mitigating urban inundation under different scenarios and hazard levels. We found that the coupled model could be applied to evaluate the effectiveness of LID for urban inundation risk mitigation. The model showed that PP were more effective for urban inundation mitigation than GR. This conclusion may be different in other regions, but it can be used by policy makers on a local basis. LID practices can only affect the inundation depth and downgrade the inundation hazard level, but cannot resolve inundation problems at a high hazard level. Therefore, other methods of stormwater management should also be applied to manage severe waterlogging. In the construction of Sponge City, people paid more attention to the effectiveness, while ignored the cost of LID. Through the analysis of cost-effectiveness, we found that wider implementation of LID practices may not lead to higher efficiency in urban watershed, and the cost and effectiveness of LID practices should be considered in the construction of Sponge City.

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Figure 1: Location and land use map of the study area in the Guangming New District of Shenzhen, China.
Figure 2: Altitude (a) and SWMM model (b) of the study area.

Figure 3: Rainfall intensity for events on 11 May 2014 and 10 May 2016 in the study area.
Figure 4: Inundation depth maps of the study area under different scenarios.
Figure 5: Reduction rates of average inundation depth (a) and inundation areas (b) under different scenarios and hazard levels.

Table 1: LID parameters in SWMM.

| LID types | structure | parameter                        | value   |
|-----------|-----------|----------------------------------|---------|
| PP        | Surface   | Berm height (mm)                 | 2       |
|           |           | Vegetation volume fraction       | 0       |
|           |           | Surface roughness (Manning’s n)  | 0.014   |
|           |           | Surface slope (%)                | 1       |
|           | Pavement  | Thickness (mm)                   | 100     |
|           |           | Void ratio (voids/solids)        | 0.25    |
|           |           | Impervious surface fraction      | 0       |
|           |           | Permeability (mm/h)              | 250     |
|           |           | Clogging factor                  | 0       |
|           | Storage   | Thickness (mm)                   | 150     |
|           |           | Void ratio (voids/solids)        | 0.4     |
|           |           | Seepage fate (mm/h)              | 1.2     |
|           |           | Clogging factor                  | 0       |
| GR        | Surface   | Berm height (mm)                 | 3       |
|           |           | Vegetation volume fraction       | 0.1     |
|           |           | Surface roughness (Manning’s n)  | 0.017   |
|           |           | Surface slope (%)                | 1       |
|           | Soil      | Thickness (mm)                   | 100     |
Porosity (volume fraction) | 0.5
---|---
Field capacity (volume fraction) | 0.2
Wilting point (volume fraction) | 0.024
Conductivity (mm/h) | 30
Conductivity slope | 5
Suction head (mm) | 60
Drainage mat | Thickness (mm) | 3
| Void fraction | 0.5
| Roughness (Manning’s n) | 0.1

**Table 2: Primary calibrated parameters in SWMM.**

| SWMM parameters | calibrated value |
|-----------------|------------------|
| N-Imperv        | 0.015            |
| N-Perv          | 0.15             |
| Dstore-Imperv/mm| 2                |
| Dstore-Perv/mm  | 5                |
| Zero-Imperv/%   | 25               |
| Roughness       | 0.013            |
| Max.Infil.Rate(mm/h) | 76            |
| Min.Infil.Rate(mm/h)  | 12             |
| Decay Constant  | 2                |
| Drying Time     | 5                |

**Table 3: Inundation depth in the observed and simulated results.**

| Inundation site | Storm on 11 May 2014 | Storm on 10 May 2016 |
|-----------------|----------------------|----------------------|
|                 | Observed  | Simulated | RE (%) | Observed  | Simulated | RE (%) |
| Gm 11           | 0.25      | 0.32      | 28      | 0.2       | 0.24      | 20     |
| Gm 12           | 0.55      | 0.69      | 25      | 0.7       | 0.75      | 7      |
| Gm 20           | 0.5       | 0.24      | -52     | 0.4       | 0.42      | 5      |
| Gm 21           | 0.45      | 0.46      | 2       | —         | —         | —      |
| Gm 24           | 0.2       | 0.26      | 30      | —         | —         | —      |
| Gm 22           | 0.2       | 0.2       | 0       | —         | —         | —      |
| Gm 16           | 0.2       | 0.23      | 15      | —         | —         | —      |

"—" means data miss, “RE” means “relative error”, unit: m.
### Table 4: Maximum inundation depth under different scenarios.

|                   | Benchmark | Scenario 1 | Scenario 2 | Scenario 3 | Scenario 4 | Scenario 5 | Scenario 6 | Scenario 7 | Scenario 8 |
|-------------------|-----------|------------|------------|------------|------------|------------|------------|------------|------------|
| maximum inundation depth (m) | 0.69      | 0.67       | 0.64       | 0.58       | 0.54       | 0.51       | 0.49       | 0.59       | 0.59       |
| Reduction rate (%)  | —         | 3          | 7          | 16         | 22         | 26         | 29         | 14         | 14         |

### Table 5: Inundation time under different scenarios and hazard levels.

|                   | Benchmark | Scenario 1 | Scenario 2 | Scenario 3 | Scenario 4 | Scenario 5 | Scenario 6 | Scenario 7 | Scenario 8 |
|-------------------|-----------|------------|------------|------------|------------|------------|------------|------------|------------|
| Low (h)           | 3.4       | 3.4        | 3.5        | 3.7        | 3.3        | 2.5        | 2.2        | 3.3        | 3.3        |
| Medium (h)        | 7.7       | 7.8        | 8          | 8.2        | 7.1        | 6          | 4.7        | 7.5        | 7.7        |
| High (h)          | 10.6      | 10.3       | 10.5       | 9.6        | 7.6        | 6          | 4.7        | 9.3        | 8.4        |
| Total (h)         | 4         | 3.9        | 4          | 4.1        | 3.6        | 2.8        | 2.3        | 3.6        | 3.6        |

### Table 6: Sensitivity of inundation to LID parameters.

| Parameter        | Inundation reduction (%) |
|------------------|--------------------------|
|                  | Permeable pavement | Green roof  |
| Berm height      | -3%                     | -1%         |
| Thickness Pavement | -15%                  | —           |
| Permeability     | -2%                     | —           |
| Thickness Soil   | —                       | -17%        |
| Porosity Soil    | —                       | -19%        |
| Conductivity Soil| —                       | 0%          |
| Thickness storage| -16%                   | —           |
| Void ratio       | -18%                   | —           |

### Table 7: CEI under different scenarios.

|                   | Scenario 1 | Scenario 2 | Scenario 3 | Scenario 4 | Scenario 5 | Scenario 6 |
|-------------------|------------|------------|------------|------------|------------|------------|
| maximum inundation depth | 0.58      | 0.72       | 0.64       | 0.44       | 0.35       | 0.29       |
| average inundation depth | 2.76      | 3.31       | 2.24       | 1.40       | 1.01       | 0.77       |
| average inundation areas  | 1.40      | 1.46       | 1.35       | 0.93       | 0.71       | 0.55       |