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A hitchhiker guide to manta rays: Patterns of association between *Mobula alfredi*, *M. birostris*, their symbionts, and other fishes in the Maldives

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Abstract

Despite being among the largest and most charismatic species in the marine environment, considerable gaps remain in our understanding of the behavioural ecology of manta rays (*Mobula alfredi, M. birostris*). Manta rays are often sighted in association with an array of smaller hitchhiker fish species, which utilise their hosts as a sanctuary for shelter, protection, and the sustenance they provide. Species interactions, rather than the species at the individual level, determine the ecological processes that drive community dynamics, support biodiversity and ecosystem health. Thus, understanding the associations within marine communities is critical to implementing effective conservation and management. However, the underlying patterns between manta rays, their symbionts, and other hitchhiker species remain elusive. Here, we explore the spatial and temporal variation in hitchhiker presence with *M. alfredi* and *M. birostris* through the Maldives and investigate the factors which may influence association using generalised linear mixed effects models (GLMM). For the first time, associations between *M. alfredi* and *M. birostris* with hitchhiker species other than those belonging to the family Echeneidae are described. A variation in the species of hitchhiker associated with *M. alfredi* and *M. birostris* was identified, with sharksucker remora (*Echeneis naucrates*) and giant remora (*Remora remora*) being the most common, respectively. Spatiotemporal variation in the presence of manta rays was identified as a driver for the occurrence of ephemeral hitchhiker associations. Near-term pregnant female *M. alfredi*, and *M. alfredi* at cleaning stations, had the highest likelihood of an association with adult *E. naucrates*. Juvenile *E. naucrates* were more likely to be associated with juvenile *M. alfredi*, and a seasonal trend in *E. naucrates* host association was identified. *Remora* were most likely to be present with female *M. birostris*, and a mean number of $1.5 \pm 0.5$ *R. remora* were observed per *M. birostris*. It is hoped these initial findings will serve as the basis for future work into the complex relationships between manta rays and their hitchhikers.
**Introduction**

Symbiosis, when considered biologically, describes a physically close and long-term association between two different species [1–3]. Symbiotic interactions are common in marine ecosystems and are fundamental in regulating the distribution, abundance, and diversity of many taxa [4, 5]. Algae-coral, anemonefish, and cleaner-client mutualisms all provide traditional examples [6–9], where at least one of the interacting species is obligately dependant on the association for all, or part, of its life-history [1, 10]. While some interactions have resulted in significant behavioural adaptions and coevolution, the competitive life of a marine species can encourage short-term and opportunistic associations in order to gain food or protection [11–15]. For example, species of the family Carangidae have been observed to associate with scalloped hammerheads (*Sphyrna lewini*) to get closer to prey items, and when following cownose rays (*Rhinoptera bonasus*), cobia (*Rachycentron canadum*) have been observed to occupy a position above their host to forage on prey rejected by the rays [16]. Pilot fish (*Naucrates doctor*) are known to commonly associate with large-bodied vertebrates such as sharks, rays and turtles [17], presumably for protection from predation [14].

Species engage in associations that vary in all degrees of intimacy, ranging from obligate to facultative, mutualistic to parasitic, and long-lived to ephemeral [1, 10, 11, 18]. These interactions, rather than the species at the individual level, determine the ecological processes that drive community dynamics, support biodiversity and ecosystem health [19]. Thus, species should not be considered in isolation, and understanding the associations within marine communities is critical to implementing effective conservation and management [5, 18, 20]. Studies that incorporate habitat-specific interactions provide an opportunity to unveil population-wide and long-term patterns into the spatial and behavioural ecology of marine fauna [20–23]. However, our understanding of marine symbionts remains limited due to the logistical challenges associated with studying complex associations in mobile organisms over large spatial scales [21, 24, 25].

Manta rays (*Mobula alfredi, M. birostris*) are large, filter-feeding batoid rays, with a pelagic existence. *Mobula birostris* has a circumglobal distribution, while *M. alfredi* has a semi-circumglobal distribution; both in tropical and subtropical waters [26–28]. They are characteristically slow to mature, have low fecundity, and exhibit migratory and aggregatory behaviours, rendering them significantly vulnerable to exploitation [26, 27]. Consequently, and because of targeted and bycatch fisheries, *M. alfredi* and *M. birostris* are classified as Vulnerable and Endangered on the IUCN’s Red List of Threatened Species, respectively [29–31]. Therefore, successful conservation of these species depends upon bridging knowledge gaps in their biology and ecology [28, 32].

The Maldives archipelago supports globally significant populations of both species of manta ray [27]. Here, coastal reef manta rays (*M. alfredi*) are commonly found throughout the archipelago, where they migrate east to west through the atolls during the transition into the Northeast (NE) Monsoon (December–March), and west to east during the onset of the Southwest (SW) Monsoon (April–November) [33]. These biannual seasonal migrations determine changes in aggregation site use, as well as the predominant behavioural activities exhibited by the highly philopatric *M. alfredi* [33, 34]. Unlike the local patterns of residency exhibited by *M. alfredi*, oceanic manta rays (*M. birostris*) are only sighted with regularity in the Maldives’ southernmost atolls of Addu and Fuvahmulah, and only for a few months each year (March–April) during the transition from the NE to the SW Monsoon [27, 35]. These southern Maldives sites are in close proximity to deep-water [27]; habitat where this species is often encountered throughout its range [36, 37]. Re-sighting rates of individuals remain extremely low, which, combined with the seasonality of sightings, suggests a large transient population which
utilises habitat away from the reef systems of the Maldives. Where the Maldives *M. birostris* originate from, or travel to, remains unknown [35].

Manta rays are often observed in association with hitchhiker fish species, such as the golden trevally (*Gnathanodon speciosus*) and members of the remora family (Echeneidae) that closely follow (within 1 m) or attach themselves to their manta ray host [15, 23]. It has been suggested that hitchhiking behaviour evolved as a means to gain protection from predation, enhance foraging opportunities, increase locomotor efficiency, and increase encounters with conspecifics [15, 17, 38–42]. However, investigations into these hitchhiker associations are limited, and the links between interspecific interactions are sensitive to the abiotic environment in which they occur [5, 43, 44].

Here, we explore the spatial and temporal variation in the presence of hitchhikers with *M. alfredi* and *M. birostris* throughout the Maldives and investigate the factors which may influence association. This study aims to improve our ecological understanding of interactions between manta rays and their hitchhikers by highlighting how these associations are structured, and what the drivers of the associations might be [11].

**Materials and methods**

**Study area**

The Maldives archipelago is comprised of 26 geographical coral atolls and approximately 2,000 islands situated predominantly in the northern Indian Ocean (Fig 1) [45]. The research was carried out under permit from the Maldives’ government (annually renewable permit: PA/2020/PSR-M07).

**Manta ray sightings and hitchhiker species**

The unique ventral body pigmentation of each manta ray enables individuals to be distinguished from one-another using a photo-identification (photo-ID) catalogue of the ventral surface of the rays [26, 27, 46]. A manta ray sighting was defined as a confirmed photo-ID of an individual manta ray on a given day at a specific location. Surveys were performed via SCUBA or freediving by trained Manta Trust staff (www.mantatrust.org) and citizen science contributors between 1987–2019. Surveys were carried out across the whole archipelago throughout the year, of all study years, although known *M. alfredi* and *M. birostris* aggregation sites were surveyed most frequently, creating some sampling bias.

Where possible each manta ray was identified to species [47], and the sex and maturity status (adult, subadult, juvenile) of each was recorded during the dive/snorkel. For each sighting, the pregnancy status (an estimate of trimester) and primary behavioural activity (cleaning, feeding, courtship, cruising, or breaching) were also recorded [27, 48]. Thereafter, two-step verification and further sighting details were determined through assessment of the image/s by trained staff using the methods described in Stevens [27], and Peel et al. [49]. For this study, manta ray sightings were removed from the analysis if the sex or maturity status of the ray could not be determined.

The presence, number, and species of hitchhikers associated with each manta ray sighting were recorded during the dive/snorkel, and verified by visual analysis of all photo-ID images (Fig 2), using FishBase to determine species identification [50]. The conspicuous colour, patterns and behaviour of cleaner fish (Labridae) enabled them to be clearly distinguished from the hitchhiker species [51]. Identified sharksucker remora (*Echeneis naucrates*) were further classified as either adults (>20 cm) or juveniles (≤20 cm) based on visual estimates against the host size [39], as well as differences between their colour and
Fig 1. A map of the Maldives archipelago located to the southwest of India. Diagram shows the 26 geographical atolls illustrated in green.

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body patterns (Fig 2). Any hitchhiker species which could not be identified (due to poor image quality), were removed from the analysis. Each site utilised by *M. alfredi* was classified by site function (feeding area, cleaning station or cruising area) based on the predominant behaviour observed at the location [34]. Almost all *M. birostris* observed were cruising, therefore no site function was investigated.
Data analysis

**Variations in hitchhiker observations.** Manta ray (*M. alfredi* and *M. birostris*) sightings and the total number of sightings where associated hitchhiker species were observed were summarised. To investigate variations in the total number of the most frequently observed hitchhiker species present with *M. alfredi*, the difference in the daily mean number of each hitchhiker species (number of sightings / number of hitchhiker species observed) was compared among manta ray sex and pregnancy stage (male, non-pregnant female, 2nd trimester pregnant female, 3rd trimester pregnant female, and 4th trimester pregnant female), season (NE or SW Monsoon), maturity status (adult, subadult, and juvenile), and site function (cleaning station, feeding area, cruising area). For *M. birostris*, categories included sex (male or female), season, and maturity status. Due to unequal sample sizes, and violation of the homogeneity of variance assumption, a Welch’s ANOVA was used followed by a Games-Howell post-hoc test using the ‘oneway.test’ function of the ‘Rmisc’ R package, and ‘oneway’ function of the ‘userfriendlyscience’ R package, respectively [52].

**Spatial and temporal variation in hitchhiker presence.** Spatial variation in the presence of the most frequently observed hitchhiker species with *M. alfredi* (adult and juvenile *E. nau-crates*) were investigated by mapping the percentage of sightings at each site (grouped by site function) where the hitchhiker species was present (total number of sightings where hitchhikers were observed / total number of sightings at the site) in ArcGIS 10.7. Any sites with a total of nine or fewer sightings (213 sites) were excluded to reduce the bias a low number of sightings may have on analysis.

Temporal variation in the presence of adult and juvenile *E. nau-crates* with *M. alfredi* was investigated using monthly time series. This series incorporated the period with the greatest number of sightings (2008–2019) to provide a suitable period from which to visualise trends (i.e., seasonality). The monthly total number of sightings were corrected for survey effort by calculating the mean monthly number of manta rays observed per survey (monthly total manta ray sightings / monthly total number of surveys).

**Generalised linear mixed models.** Logistic generalised linear mixed models (GLMM) using R v4.0.0 [52] were used to investigate relationships between the presence of the most frequently observed hitchhiker species (adult and juvenile *E. nau-crates*, *G. speciosus*, and *Lutjanus bohar*) with *M. alfredi* and four explanatory variables: sex with pregnancy status, maturity status, site function (determined by the predominant behaviour observed at the site [34]), and seasonality (NE or SW Monsoon). Due to the low number of recorded associations between *M. alfredi* and most of the hitchhiker species, only those with sufficient data were included in the GLMM analysis. The same model was used for *Remora remora* (the most frequently observed hitchhiker species with *M. birostris*), but without site function, and sex was classified only as male or female as pregnancies were only observed during four sightings. Each GLMM was fitted with a logit link function to the binary response of hitchhiker species presence (1) and absence (0) using the ‘lme4’ R package [53]. Each model contained the manta-ID as a random intercept to account for any temporal autocorrelation arising from individual rays being repeatedly observed [54]. To compare the relative goodness-of-fit, GLMM models without random effects (GLM) were tested. To reliably estimate the parameters, categories of variables with levels observed equal to or less than five times were removed. For example, under the category behavioural activity, the level ‘breaching’ was observed on less than five occasions, so was removed from analysis. The most informative explanatory variables were identified by firstly testing GLMM models with all combinations of explanatory variables. The variance inflation factor (VIF) was used to test models for multicollinearity; the maximum VIF was <1.5. Model performance was assessed using corrected Akaike information criterion (AICc).
test statistic [55] using the ‘MuMin’ R package [56], and the DHARMa R package [57] was used to check the model residuals were normally distributed. The highest-ranking models (with the lowest AICc value, S1 Table) for each hitchhiker species were then interpreted in terms of odds ratios (ORs) (the likelihood of the presence of the hitchhiker species in comparison with the reference category). Any models with ΔAICc <2 were considered in interpretation of the highest-ranking model [55]. The significance of each explanatory variable was determined by the 95% confidence interval (CI) of OR, whereby a narrower CI indicates a more precise estimation while, in comparison, a wider CI which had a greater uncertainty. A CI that crossed one is considered non-significant. Any ORs with \( p > 0.05 \) are not reported.

**Results**

**Reef manta ray (M. alfredi) sightings and associated hitchhiker species (1987–2019)**

A total of 4901 *M. alfredi* were individually identified [male = 2442 (50%), female = 2459 (50%)] during a total of 72912 sightings, of which 44071 (60%) were of females [adult = 25700 (58%), juvenile = 18371 (42%)] and 28841 (40%) were males [adult = 25968 (90%), sub-adult = 1443 (5%), juvenile = 1430 (5%)]. All sightings occurred across 353 sites, of which 95 (27%) were cleaning stations [sightings = 24034 (33%)], 53 (15%) were cruising areas [sightings = 129 (0%)], and 205 (58%) feeding areas [sightings = 48749 (67%)].

Twelve different species of hitchhiker were identified with *M. alfredi* (Table 1 and Fig 3). The most frequently observed hitchhiker species with *M. alfredi* was *E. naucrates*. Adult *E.*

| Manta ray species | Hitchhiker species                  | Total no. hitchhiker individuals observed | No. ray sightings when observed | Max no. observed per sighting | Mean no. per sighting ±SD |
|-------------------|-------------------------------------|-------------------------------------------|--------------------------------|-------------------------------|--------------------------|
| *Mobula alfredi*  | Caranx melampygus                   | 53                                        | 44                             | 9                             | 1.2 ± 1.2                |
|                   | Caranx ignobilis                    | 26                                        | 26                             | 1                             | 1 ± 0                    |
|                   | Gnathanodon speciosus               | 1176                                       | 536                            | 29                            | 2.1 ± 3.2                |
|                   | Naucrates doctor                    | 9                                         | 6                              | 3                             | 1.5 ± 0.8                |
|                   | Elagatis bipinnulata                | 28                                        | 25                             | 3                             | 1.2 ± 0.4                |
|                   | Echeneis naucrates                  | 16549                                     | 8211                           | 24                            | 2 ± 1.4                  |
|                   | *Echeneis naucrates* (juv.)         | 1025                                      | 967                            | 4                             | 1.1 ± 0.3                |
|                   | Remora remora                       | 1                                         | 1                              | 1                             | 1                        |
|                   | Remora albescens                    | 41                                        | 40                             | 2                             | 1 ± 0.2                  |
|                   | Rachycentron canadum                | 18                                        | 18                             | 1                             | 1 ± 0                    |
|                   | Lutjanus bohar                      | 247                                       | 228                            | 3                             | 1.1 ± 0.3                |
|                   | Aulostomus chinensis                | 3                                         | 3                              | 1                             | 1 ± 0                    |
| *Mobula birostris*| Caranx lugubris                     | 31                                        | 25                             | 3                             | 1.2 ± 0.5                |
|                   | Naucrates doctor                    | 9                                         | 1                              | 9                             | 9                        |
|                   | *Echeneis naucrates* (juv.)         | 6                                         | 2                              | 3                             | 3 ± 0                    |
|                   | *Echeneis naucrates* (juv.)         | 2                                         | 2                              | 1                             | 1 ± 0                    |
|                   | Remora remora                       | 582                                       | 398                            | 3                             | 1.5 ± 0.5                |

Total and mean number of hitchhikers observed with *Mobula alfredi* between 1987–2019 (total sightings = 72912) and *Mobula birostris* between 1996–2019 (total sightings = 726).

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naucrates were observed with M. alfredi during 7189 (10%) of the total sightings, and juveniles during 756 (1%) sightings.

**Variations in hitchhiker associations with reef manta rays (M. alfredi)**

Due to the low number of recorded associations between M. alfredi and most of the hitchhiker species, only E. naucrates was investigated further here.

**Adult sharksucker remora (E. naucrates).** When present, the number of adult E. naucrates associated with M. alfredi ranged between one and twenty-four individuals per sighting. There was a significant difference in the daily mean number of adult E. naucrates associated with M. alfredi between manta ray sex and pregnancy stage ($F_{4, 1302} = 102.6, p < 0.001$). The highest mean number of E. naucrates occurred with female M. alfredi in their 4th trimester of pregnancy; significantly higher ($p < 0.001$) than with males, non-pregnant females, and 2nd trimester pregnant females (Fig 4). There was also a significant difference between maturity status categories ($F_{2, 3280} = 123.1, p < 0.001$), where the highest mean number of E. naucrates were observed with adults, which was significantly higher ($p < 0.001$) than with subadults and juveniles (Fig 4). The total mean number of adult E. naucrates with M. alfredi was also significantly different between site functions ($F_{2, 1223} = 222.1, p < 0.001$), which was significantly higher ($p < 0.001$) at cleaning stations and cruising areas than at feeding areas (Fig 4). The total mean number of adult E. naucrates associated with M. alfredi was also significantly different during each season ($F_{1, 6431} = 155.7, p < 0.001$), with more E. naucrates associated with M. alfredi during the NE Monsoon ($p < 0.001$) (Fig 4).
Juvenile sharksucker remora (*E. naucrates*). When present, the number of juvenile *E. naucrates* associated with *M. alfredi* ranged between one and four individuals per sighting. There was a significant difference in the daily mean number of juvenile *E. naucrates* observed between manta ray sex and pregnancy stage (*F*₄, 1496 = 19.7, *p* < 0.001). The highest daily mean number of juveniles occurred with male *M. alfredi*, which was significantly higher (*p* < 0.001) than pregnant females (Fig 5). There was a significant difference in the maturity status category (*F*₂, 2661 = 15.6, *p* < 0.001), where the highest mean number of juvenile *E. naucrates* were observed with juvenile *M. alfredi*; significantly higher than with adults (*p* < 0.001). The mean number of juvenile *E. naucrates* associated with *M. alfredi* was also significantly different between site functions (*F*₂, 1204 = 30.3, *p* < 0.001), with a significantly higher (*p* < 0.001) amount at feeding areas than at cleaning stations. The daily mean number of juvenile *E. naucrates* associated with *M. alfredi* was also significantly different during each season (*F*₁, 5828 = 19.3, *p* < 0.001), with a significantly higher number of *E. naucrates* associated with *M. alfredi* during the NE Monsoon (*p* < 0.001).

**Spatial and temporal variation in hitchhiker presence.** There was a total of 72361 sightings across 149 sites which had ten or more manta ray sightings; 23926 (33%) occurred at cleaning stations where *E. naucrates* were present during 4474 (19%) of observations, and 48435 (67%) occurred at feeding areas where *E. naucrates* were present during 2641 (5%) of observations. The highest percentage of *E. naucrates* presence by atoll occurred within Meemu (40%), North Malé (27%), and Ari (27%) Atolls. The sites with the highest percentage of *E. naucrates* present were Maayafushi Fallhu (60%), a feeding area in Ari Atoll, at Rangali Madi-varu (50%), a cleaning station in Ari Atoll, and at Delidhoo (50%), a cleaning station in Thiladhunmathi Atoll. *Echeneis naucrates* were not observed with *M. alfredi* at 25 sites, 17 of which are feeding areas (total sightings = 1512) and 8 were cleaning stations (total sightings = 266).
The mean percentage of \( E. \) naucrates present by site function was 11.6±12.2% at feeding areas, and 12.8±14.4% at cleaning stations (Fig 6).

The proportion of adult \( E. \) naucrates observed with \( M. \) alfredi was highest during the NE Monsoon, typically from January to March, at which time the lowest monthly mean number of \( M. \) alfredi sightings occurred (Fig 7). There was no seasonal pattern observed in the presence of juvenile \( E. \) naucrates with \( M. \) alfredi.

Reef manta ray (\( M. \) alfredi) generalised linear mixed models

**Adult sharksucker remora (\( E. \) naucrates).** The highest ranking GLMM for \( E. \) naucrates contained all four predictors (site function, sex, maturity, and season) (S1 Table). The model (Fig 8) suggests that \( E. \) naucrates were most likely to be present with \( M. \) alfredi at cleaning stations, which were 49% more likely than at feeding areas (OR = 0.51). This hitchhiker species was also most likely to be present on females in their 4\(^{th}\) trimester of pregnancy (OR = 2.6), which was 160% higher than non-pregnant females (reference category), while males (OR = 0.59) were 41% less likely to have \( E. \) naucrates hitchhikers than non-pregnant females. \( Echeneis \) naucrates were least likely to be present on juvenile \( M. \) alfredi (OR = 0.67), which were 31% less likely to have this hitchhiker species than adults (reference category). The GLMM also indicates that the likelihood of \( E. \) naucrates presence with \( M. \) alfredi was highest during the NE Monsoon (reference category), which was 25% more likely than during the SW Monsoon (OR = 0.75).

**Juvenile sharksucker remora (\( E. \) naucrates).** The highest ranking GLMM for juvenile \( E. \) naucrates contained all four predictors (site function, sex, maturity, and season). The model (Fig 8) suggests that juvenile \( E. \) naucrates were most likely to be present with \( M. \) alfredi at feeding areas (OR = 2.08), which was 108% more likely than at cleaning stations (reference
category). Juvenile *E. naucrates* were also more likely to be present on juvenile *M. alfredi* (OR = 2.37), which were 137% more likely to have this hitchhiker group than adults (reference category). Male *M. alfredi* (OR = 1.59) were 59% more likely than females to have juvenile *E. naucrates* present, and juvenile *E. naucrates* were least likely to be present with *M. alfredi* during the SW Monsoon (OR = 0.63); 37% less likely than during the NE Monsoon (reference category).

**Red snapper (*L. bohar*)**. The highest ranking GLMM for *L. bohar* (Fig 8) contained one significant predictor (site function) and suggests that the species were most likely to be present with *M. alfredi* at cleaning stations (reference category), which was 100% more likely than at feeding areas (OR = 0). There was one model with ΔAICc < 2 (S1 Table), which was the same

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**Fig 6. Heatmaps coloured by season and percentage of sightings where *Echeneis naucrates* were present.** Includes feeding areas and cleaning stations with > 10 *E. naucrates* sightings.

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**Fig 7. Time series plot showing *Mobula alfredi* sightings and *Echeneis naucrates* presence.** The total monthly number of *Mobula alfredi* sightings 2008–2019, and the percentage of those that had *Echeneis naucrates* associations.

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as the highest-ranking model with the addition of season, but this predictor was non-significant ($p > 0.05$).

**Golden trevally** *(G. speciosus)*. The highest-ranking GLMM for *G. speciosus* contained two significant predictors (site function and maturity status). *Gnathanodon speciosus* were most likely to be present with *M. alfredi* at cleaning stations (reference category), which was 22% more likely than at feeding areas ($OR = 0.78$) (Fig 8). There were two models with $\Delta AICc < 2$ (S1 Table). One of these models contained the same predictors as the highest-ranking model with the addition of season, but this predictor was non-significant ($p > 0.05$). The other model contained only the predictor maturity status, which suggested *G. speciosus* were more likely to be present on juvenile *M. alfredi* ($OR = 1.5$), which were 55% more likely to have this hitchhiker species than adults (reference category).

**Oceanic manta ray** *(M. birostris)* sightings and associated hitchhiker species

A total of 663 *M. birostris* were individually identified [male = 363 (55%), female = 300 (45%)] during a total of 726 sightings, of which 329 were females [adult = 237 (72%), subadult = 24
and 397 were males [adult = 371 (93%), subadult = 24 (6%), juvenile = 2 (1%)]. All sightings occurred across 39 sites, of which, 642 (88%) occurred at Fuvahmulah Atoll, and 662 (91%) occurred during the months of March and April, straddling the transition between the NE and SW Monsoons. *Mobula birostris* were observed exhibiting cleaning [sightings = 8 (1%)], cruising [sightings = 681 (94%)], feeding [sightings = 10 (1%)], and courtship behaviour [sightings = 27 (4%)].

Five different hitchhiker species were identified associated with *M. birostris* (*Table 1* and *Fig 3*). The most frequently observed hitchhiker species was *R. remora*, which was observed with *M. birostris* during 397 (55%) sightings.

**Variations in hitchhiker associations with oceanic manta rays (*M. birostris*)**. Due to the low number of recorded associations between *M. birostris* and most of the hitchhiker species, only *R. remora* was investigated further here.

When present, the number of *R. remora* associated with *M. birostris* ranged between one and three individuals per sighting. The highest daily mean number of *R. remora* occurred with female *M. birostris*, which was significantly higher than males (*F*₁, 233 = 12.7, *p* < 0.001). There were no significant differences for *M. birostris* maturity status, or between seasons (*Fig 9*).

**Oceanic manta ray (*M. birostris*) generalised linear mixed models.** *Remora remora* were most likely to be present with female *M. birostris* (reference category); 41% (OR = 0.59) more likely than with males (*Fig 8*), and they were more likely to be present during the SW Monsoon which was 49% higher (OR = 1.49) than during the NE Monsoon (reference category) (*Fig 8*). There were two models with ΔAICc < 2 (*S1 Table*). Both models contained the same predictors as the highest-ranking model (sex and season) with the addition of maturity status. This GLMM model suggests *R. remora* may also be most likely to be present with adult *M. birostris* (reference category); 47% (OR = 0.53) more likely than with juveniles.
Discussion

A variation in the species of hitchhiker associated with *M. alfredi* and *M. birostris* was identified, with *E. naucrates* and *R. remora* being the most common, respectively. Spatiotemporal variation in the presence of manta rays was identified as a driver for the occurrence of ephemeral hitchhiker associations, and for the first time, associations between *M. alfredi* and *M. birostris* with hitchhiker species other than those belonging to the family Echeneidae are described.

Reef manta ray (*M. alfredi*)

Twelve hitchhiker species were observed associating with *M. alfredi*, of these, *E. naucrates* was by far the most frequent. The Echeneidae family are well-known for their hitchhiking behaviour on large-bodied vertebrates [23, 39, 41]. The relationship with their hosts is generally considered mutualistically symbiotic, as most remora species spend all their post-larval life in close association with their hosts, eating their ectoparasites in return for a range of benefits [15, 39, 40, 58]. However, the degree of host specificity and the nature of the association has been shown to vary along the symbiotic continuum (mutualism, commensalism, and parasitism), and the importance of this host food source varies between remora species and at different life stages [1, 25, 38, 39]. Recent investigations into the echeneid-host association also suggest that there may be significant costs incurred by the host, such as persistent damage and scarring from the adhesive disc by which the remoras attach themselves to their hosts, as well as deforming injuries when smaller-sized remoras force themselves through the gill slits and other body openings of their mobulid host [15, 23, 59].

*Echeneis naucrates* is a neritic species that is presumed to be physiologically unable to remain attached to their host when they dive at depth [15, 22, 58]. *Mobula alfredi* in the Maldives and elsewhere throughout their range are known to undertake regular dives below 200 m, presumably to forage on prey within the deep-scattering layer [33, 60–63]. Consequently, we suggest *E. naucrates* associations with their *M. alfredi* hosts are often ephemeral, with adults spending significant periods of time free-swimming, re-associating with a host upon their return to the remora’s habitat [15, 58]. Given other hosts of adult *E. naucrates* are also known to make regular dives below the Neritic Zone [64, 65], it is likely associations with these hosts are also ephemeral. Indeed, it is unknown what percentage of their adult lives *E. naucrates* spend away from their hosts, but it is not uncommon for this species to be observed free-swimming, or resting on the seabed in groups, during reef dives in the Maldives (Stevens, pers. obs.).

*Echeneis naucrates* associations with *M. alfredi* varied significantly depending on host sex, pregnancy state, and maturity status. Associations also varied significantly depending on the function of the observation site and the season within which the sighting occurred. *Echeneis naucrates* were significantly more likely to be associated with female *M. alfredi* than males, and with *M. alfredi* at cleaning stations. Previous studies have demonstrated that female *M. alfredi* are significantly more likely to be sighted at cleaning stations than males [27, 66]. Cleaning stations are predominantly located on shallow reefs or in lagoons [27]; suitable habitat for the neritic *E. naucrates* [58]. Therefore, more frequent utilisation of cleaning stations by female *M. alfredi* provides greater opportunity for an association to occur, and the more time spent at these sites, the greater the opportunity for a higher total number of associations to occur. In contrast, many shallow *M. alfredi* feeding sites in the Maldives, and elsewhere, are situated in atoll channels which are adjacent to deep-water areas, where it is hypothesised *M. alfredi* forage prior to shallow-water feeding events [67]. As *E. naucrates* may be physiologically unable to remain associated with *M. alfredi* when they travel to these deep-water locations, the likelihood of association during subsequent shallow-water feeding is reduced [15, 67].
Near-term pregnant female *M. alfredi* had the highest likelihood of an association, and the highest mean number of *E. naucrates* associates per individual; consistent with a previous preliminary study in the region [68]. The study suggested these increased associations were a result of thermoregulatory advantages gained by pregnant *M. alfredi* occupying warm-water habitat for longer periods during late-term gestation to reduce gestation times [15, 68], a common behaviour documented in other elasmobranchs [69–71]. If this hypothesis is correct, it would explain, in part, why female *M. alfredi* generally are recorded more frequently at cleaning stations than males. And as stated above, the longer the continuous period an individual *M. alfredi* spends in suitable *E. naucrates* habitat, the greater the chance of an association/s occurring.

A seasonal reproductive strategy, something which has been observed in both captive [72] and wild-caught *E. naucrates* [22], is typically adopted by species whose access to resources, such as prey and host availability, is uncertain; aiming to maximise juvenile recruitment [22, 73]. Bachman et al. [22] highlighted that hitchhiker behaviour in terrestrial arthropods predicts that host association restricts mate selection, reproduction, and location. Therefore, eche- neid population dynamics are likely to be strongly influenced by the ecology and availability of their hosts. For juvenile *E. naucrates*, host parasitic copepods comprise a more integral part of their diet than that of an adult [38], and juvenile remoras that are not attached to a host may be exposed to unsuitable environments and increased predation risk [58]. During the current study, juvenile *E. naucrates* were more likely to be associated with juvenile *M. alfredi*, most likely because juvenile manta rays spend most of their time in protected lagoons and other shallow water nursery habitats [27, 66], increasing the chance of long-term associations between the two species for reasons already discussed in this study. Thus, juvenile manta rays are likely to provide a more suitable host for the juvenile remoras that require continual host associations in shallow water to survive to adulthood [22, 27, 58]. The obligate symbiosis of juvenile *E. naucrates* with their hosts may then transition to a more facultative relationship in adult remoras [22, 58].

Associations between *E. naucrates* and *M. alfredi* were also significantly higher during the NE Monsoon; the first record of a seasonal trend in *E. naucrates* host association. The higher association rate during the NE Monsoon may potentially be associated with the suppression of primary productivity, which peaks during the Maldives’ SW Monsoon [33, 34]. As proposed sites of behavioural thermoregulation and predator avoidance [15, 27], cleaning stations may be utilised more during periods of lower primary productivity to conserve energy and reduce risk of predation. Thus, the greater period a manta ray spends within such habitat, the greater chance of an association occurring with a *E. naucrates*. However, greater abundances of large skipjack (*Katsuwonus pelamis*) and yellowfin tuna (*Thunnus albacares*) have been observed in the Maldives in the less productive NE Monsoon [33]. This led Anderson et al. [33] to suggest that there may be an increase in primary productivity at this time associated with a deep chlorophyll maximum (DCM) that is not visible to the satellite technology (SeaWiFS) which records the high SW Monsoon productivity. Manta rays are poikilothermic, with an optimal thermal temperature of 20–26˚C, but can endure colder temperatures for short periods due to a counter-current heat-exchange mechanisms [63, 74]. This physiological adaption enables manta rays to forage on zooplankton blooms within the deep-scattering layer, and down to depths of over 672 metres and temperatures of 7.6˚C [63]. If Anderson et al. [33] hypothesis is correct, basking in warmer shallower waters during the NE Monsoon, at sites like cleaning stations, prior to and post deep forays may enable manta rays to physiologically prepare for, and recover from, the large metabolic costs incurred from such deep foraging bouts [63, 75]. This behavioural thermoregulation has also been used to explain why the spinetail devil ray (*Mobula mobular*) [76], the whale shark (*Rhincodon typus*) [75], and tuna species return to
shallow habitats after deep dives [77]. To support this hypothesis, more data on the diving behaviour and habitat use of manta ray in the Maldives is required, utilising camera-mounted and satellite tracking technologies, along with chlorophyll-α measurements, to address this question.

Considering the ecology of both the host and the symbiont, the results of this study suggest that the patterns of association between *E. naucrates* and *M. alfredi* are most likely driven by the spatiotemporal variation in presence of manta rays in the sharksucker’s habitat. The overlap in species habitat-use at cleaning sites and within sheltered lagoons, particularly during the NE Monsoon, provides an explanation for why near-term pregnant female *M. alfredi* have the greatest likelihood of associating with adult *E. naucrates*, and why juvenile *M. alfredi* have the greatest likelihood of associating with juvenile *E. naucrates*.

Here, for the first time, associations between *M. alfredi* and hitchhiker species, other than those belonging to the family Echeneidae, are described. In reef fish systems, follower-feeding associations are influential on the structure of surrounding reef communities [11]. Competitive life in reef habitats brings about short-term associations to obtain food [11–13]. However, many ephemeral interactions between small teleost and megafauna species, such as the one described here in the introduction between *R. canadum* and *R. bonasus* [16], are yet to be investigated. Considering the carnivorous feeding ecology of many of the non-echeneid hitchhikers identified here, such as the black (*Caranx lugubris*), bluefin (*C. melampygus*) [78], and giant (*C. ignobilis*) trevallies [79], as well as the rainbow runner (*Elagatis bipinnulata*) [80], cobia (*Rachycentron canadum*) [16], red snapper (*Lutjanus bohar*) [81], and Chinese trumpetfish (*Aulostomus chinensis*) [82], it is likely these associates also opportunistically utilise the body of the manta ray to get near their prey [11–13, 82]. An observation supported by the authors in the field during this study for all the aforementioned species.

Several of the other hitchhiker associations with *M. alfredi*, such as the golden trevally (*Gnathanodon speciosus*) and the pilot fish (*Naucrates doctor*), are likely to be driven primarily by the advantage of the shelter provided by the host [14, 15]. In the case of *G. speciosus*, these associations only last until the juveniles are large enough to survive by themselves [15, 83]. Aside from *N. doctor* (which was only sighted with *M. alfredi* on six occasions during the study), all the non-remora hitchhiker species identified are neritic reef-dwellers [17, 84]. Therefore, there are likely to be more opportunities for associations to form between these species and *M. alfredi* than with the predominantly oceanic *M. birostris* [37]. This hypothesis is supported by the results of the GLMM models, which predicted the greatest chance of association rates between *M. alfredi*, *L. bohar* and *G. speciosus* at cleaning stations. Furthermore, as with juvenile *E. naucrates*, juvenile *G. speciosus* were significantly more likely to be associated with juvenile *M. alfredi*. The models also suggest that maturity status might be more influential than site function as one model with ΔAICc <2 contained only the predictor maturity status. It is possible the juvenile manta rays are similarly providing a more suitable host for the juvenile trevallies, which may also require continual host associations in shallow water to survive to adulthood [15, 27]. All of the non-echeneid hitchhiker species identified are considered to engage in commensalism, as they obtain benefits whilst their mobulid host receives neither benefits or harm from the association [1, 15–18, 83]. Overall, the hitchhiker species which associated most with *M. alfredi* reflect the characteristic assemblage of species that occupy the habitat where the host spends time [21].

The lower number of associations of non-echeneid hitchhiker species with *M. alfredi* could be a result of the analysis techniques used in this study, as sightings images were primarily focused on the ventral surface of the ray (to get a suitable photo-ID), but species such as *G. speciosus* often reside around the head of their host, exhibiting ‘piloting’ behaviour [15, 83]. Similar sampling bias is also likely to explain the low number of sightings where little remora
(Remora albescens) were present \( (n = 40) \) in this study [15]. This cryptic and poorly studied remora species was rarely observed outside the body of the manta rays. However, opportunistic visual inspection inside the buccal cavity of feeding adult M. alfredi by the authors during the study often revealed a pair of this species attached to the upper cavity. Therefore, most associations between these two species during this study were probably missed but were likely common.

**Oceanic manta ray (M. birostris)**

Five hitchhiker species were found to be associated with M. birostris. Of these, R. remora was by far the most frequently observed, unlike in M. alfredi, where only one association was recorded between these two species during this study. And unlike the M. alfredi and adult E. naucrates association, the symbiosis between M. birostris and R. remora appears long-term, with the remora rarely, if at all, leaving the protection of its host [15, 23, 38], even when feeding at 130–140 m depth [37]. Previous examination of the diet data revealed that parasitic copepods comprise a crucial part of the R. remora diet, but that the importance decreases as the remora increases in size [25, 38, 39]. Despite this, the obligate nature of R. remora, like that of juvenile E. naucrates [58], further suggests that echeneid population dynamics are influenced by the distribution patterns of its host [22, 58].

In Mexico’s Revillagigedo Archipelago’s National Park, where the only other peer-reviewed study of manta hitchhiker associations has been undertaken, Becerril-García et al. [23] found no association between the total number of R. remora, M. birostris sex, morphotype, and month of the year. As a result, it was suggested that the presence of remoras could be influenced by the level of host ectoparasites, population size, diving behaviour, and surrounding environmental conditions. In the present study, a mean number of 1.5 ± 0.5 R. remora were observed per M. birostris sighting. In Mexico, a mean number of 1.6 ± 0.6 R. remora attached to each manta were observed [23], in Peru there has been a sighting of eleven echeneids associated with a single M. birostris [85], and at Isla de la Plata in Ecuador, as many as 40 individual R. remora have been recorded associating with a single M. birostris (Guerrero, pers. comm.; Harty pers. comm.). Indeed, at Isla de la Plata, large numbers of R. remora are frequently associated with their manta hosts. Regional ecological variations between M. birostris populations, as well as the surrounding environmental conditions, are likely to be influencing the presence of R. remora. The significant differences in R. remora association rates recorded between male and female M. birostris in this study may also be linked to either foraging or reproductive strategies [35], although much more knowledge of M. birostris habitat use and behavioural ecology is required to address this hypothesis satisfactorily. Therefore, research into the ecological variations within and between M. birostris populations is a topic worthy of future research and could reveal valuable insight into the ecology of both the host and the symbiont [23].

Due to logistical constraints, accurate size data on the sighted manta rays during this study was rarely collected. Therefore, variation in manta ray disc width (DW) and hitchhiker presence was not investigated here. However, future directed studies could investigate DW (i.e., the sexual dimorphism between male and female manta rays) as a potential further pattern of association. Furthermore, the potential hosts costs and benefits of different echeneid densities remains relatively unexplored [59]. Research into these factors could also provide further insight into the patterns of association identified within the current study.

Unlike E. naucrates, R. remora frequently attach themselves to the dorsal surface of a manta ray (Stevens, pers. obs.). However, despite their locality, these large remoras where often still visible in the images because their heads or tails protruded from the edge of the host’s body due to their favoured attachment positioning. Nonetheless, as the photo-ID images collected
in this study were primarily focused on the ventral surface of the rays, the presence of some *R. remora* were likely to have been missed. Future studies into associations between these two species should attempt to collect both ventral and dorsal imagery to address this methodological weakness.

Despite biological associations often being one of the first components of biodiversity to be altered by abiotic change, the associations between interacting species are often overlooked in regard to our changing world [5, 43, 86]. Disconcertingly, the climatic crisis and other anthropogenic threats in the Maldives are becoming increasingly apparent, with weakening monsoon winds, rising sea surface temperatures and levels, reef degradation, overfishing and habitat destruction all effecting the resilience of the ecosystems and the life they support [45, 87–89]. Changes in the Maldives environment determines the spatial and temporal variation in the presence and behaviour of manta rays [33, 34], which in turn drives ephemeral hitchhiker associations. This is an important consideration because the fitness benefits, and the degree of dependency between hitchhikers and manta rays, remain unknown, while declines in host species may alter hitchhiker populations [22, 58]. Research has already identified the potential for a reduction in the stability and pervasiveness of cleaner-client interactions under environmental change [90]. Thus, it begs the question of how stable hitchhiker associations will be under increasingly unstable environmental conditions [44, 91]. Therefore, an enhanced effort to document and understand these symbiotic interactions is critical.

### Conclusions

Manta rays provide a midwater habitat for a broad range of species that require the protection and sustenance these hosts afford [15, 22, 23, 68]. The current study identified a variation in the species of hitchhiker associated with *M. alfredi* and *M. birostris*, with *E. naucrates* and *R. remora* being the most common, respectively. Patterns of association in the presence of a range of hitchhiker species were identified, with spatiotemporal variation in the presence of manta rays acting as a driver for the occurrence of ephemeral hitchhiker associations. Of particular interest, near-term pregnant female *M. alfredi*, and *M. alfredi* at cleaning stations had the highest likelihood of an association with adult *E. naucrates*. Juvenile *E. naucrates* were more likely to be associated with juvenile *M. alfredi*, and a seasonal trend in *E. naucrates* host association was identified. Until now, these interactions have remained undocumented or briefly addressed in the literature.

Given the rapid pace at which anthropogenic activities are altering oceans worldwide, significant effort should be aimed at understanding these associations [5, 92]. The current study intends to serve as a basis for a deeper understanding of the symbiotic relationships and other associations which occur between manta rays and their hitchhikers, which in turn we hope will ultimately elucidate our knowledge of both the host and the hosted in a more ecologically meaningful way.

Further research of hitchhikers in different manta ray populations is warranted to evaluate whether the associations and structures found within the Maldives apply to other geographic locations, as well as understanding the drivers of the association more holistically. While it could be said that these hitchhikers are just along for the ride, they could also play a valuable role in the ecological understanding and conservation of such economically valuable and vulnerable species.

### Supporting information

S1 Table. Summary of AICc relative goodness of fit metric values from the GLMM modeling procedure. Includes all combinations of explanatory variales for each hitchhiker species
identified with *Mobula alfredi* and *M. birostris*.

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**References**

1. Dimijian GG. Evolving Together: The Biology of Symbiosis, Part 1. Baylor Univ Med Cent Proc. 2000; 13. [https://doi.org/10.1080/08998280.2000.11927676](https://doi.org/10.1080/08998280.2000.11927676) PMID: 16389385

2. Dougal AE. Symbiotic interactions. Oxford: Oxford University Press; 1994. [https://doi.org/10.1259/0007-1285-67-800-813](https://doi.org/10.1259/0007-1285-67-800-813) PMID: 8087489

3. Paracer S, Ahmadjian V. Symbiosis: an introduction to biological associations. 2nd ed. New York: Oxford University Press; 2000.

4. Castro P. Animal symbioses in coral reef communities: a review. Symbiosis. 1988; 5: 161–184.

5. Auster PJ, Estes JA, Coleman FC. Species interactions in marine communities: the invisible fabric of nature. Bull Mar Sci. 2013; 89(1): 3–9. [https://doi.org/10.1073/pnas.1305853111](https://doi.org/10.1073/pnas.1305853111) PMID: 24367087

6. Sheppard C, Davy S, Pilling G, Graham N. The Biology of Coral Reefs. 2nd ed. Oxford University Press; 2018. [https://doi.org/10.1007/s00338-018-01737-w](https://doi.org/10.1007/s00338-018-01737-w) PMID: 30930680

7. Fautin D. The anemonefish symbiosis: what is known and what is not. Symbiosis. 1991; 10: 23–46.

8. Losey GS. The Ecological Importance of Cleaning Symbiosis. Copeia. 1972; 44: 820–833. [https://doi.org/10.2307/1442741](https://doi.org/10.2307/1442741)

9. Grossman A, Sazima C, Bellini C, Sazima I. Cleaning symbiosis between Hawksbill turtles and reef fishes at Fernando de Noronha Archipelago, off Northeast Brazil. Chelonian Conserv Biol. 2006; 5(2). [https://doi.org/10.2744/1071-8443(2006)5[284:CSBHTA]2.0.CO;2](https://doi.org/10.2744/1071-8443(2006)5[284:CSBHTA]2.0.CO;2)

10. Côté I. Evolution and Ecology of Cleaning Symbioses in the Sea. In: Barnes M, Gibson N, editors. Oceanography and Marine Biology: Biology: An Annual Review. CRC Press; 2000. pp. 311–355.
11. Inagaki KY, Mendes TC, Quimbayo JP, Cantor M, Sazima I. The structure of fish follower-feeding associations at three oceanic islands in southwestern Atlantic. Environ Biol Fishes. 2020; 103(12): 1–11. https://doi.org/10.1007/s10641-019-00924-0

12. Sabino J, Andrade LP, Sazima I, Teresa F, Floeter SR, Sazima C, et al. Following fish feeding associations in marine and freshwater habitats. Mar Freshw Res. 2017; 68(2): 381–387. https://doi.org/10.1080/13629904.2012.673014

13. Pereira PHC, Feitosa JLL, Chaves LCT, De Araújo ME. Reef fish foraging associations: “Nuclear-follower” behavior or an ephemeral interaction? Proc 12th Int Coral Reef Symp. 2012.

14. Magnuson JJ, Gooding RM. Color Patterns of Pilotfish (Naucrates ductor) and Their Possible Significance. Copeia. 1971; 2: 314. https://doi.org/10.2307/1442834

15. Stevens G, Fernando D, Dando M, di Sciara GN. Guide to the Manta and Devil Rays of the World. Princeton University Press; 2018. https://doi.org/10.2307/j.ctvs32s7t

16. Michael S. Reef sharks and rays of the world. A guide to their identification, behaviour, and ecology. Monterey, California: Sea Challengers.; 1993. https://doi.org/10.1071/MF15294

17. Vassilopoulos V, Siapatis A, Christides G, Bekas P. The biology and ecology of juvenile pilotfish (Naucrates ductor) associated with Fish Aggregating Devices (FADs) in eastern Mediterranean waters. 2004; 5(1) https://doi.org/10.12681/mms.211

18. Rico-Gray V. Interspecific Interaction. Encyl Life Sci. Chichester, UK: John Wiley & Sons, Ltd; 2001. https://doi.org/10.1038/ngp.eis.0003280 PMID: 2166960

19. Simmons BI, Wauchop HS, Amano T, Dicks L V., Sutherland WJ, Dakos V. Vulnerable species interactions are important for the stability of mutualistic networks. bioRxiv. 2019. https://doi.org/10.1101/604868

20. Armstrong AO, Armstrong AJ, Bennett MB, Richardson AJ, Townsend KA, Everett JD, et al. Mutualism promotes site selection in a large marine planktivore. Ecol Evol. 2021; 1–18. https://doi.org/10.1002/ece3.7464 PMID: 30198143

21. Bachman BA, Kraus R, Peterson CT, Grubbs RD, Peters EC. Growth and reproduction of Echeneis naucrates from the eastern Gulf of Mexico. J Fish Biol. 2018; 93: 755–758. https://doi.org/10.1111/jfb.13790 PMID: 30198143

22. Becerril-García EE, Gutiérrez-Ortiz MA, Preciado-González PA, Ayala-Bocos A. Presence of Remora on Mobula birostris in Revillagigedo National Park, Mexico. Mar Freshw Res. 2020; 71: 414–417. https://doi.org/10.1071/MF19089

23. Graham RT, Witt MJ, Castellanos DW, Remolina F, Maxwell S, Godley BJ, et al. Satellite tracking of manta rays highlights challenges to their conservation. PLoS One. 2012; 7(5). https://doi.org/10.1371/journal.pone.0036834 PMID: 22590622

24. Marshall A, Kashiwagi T, Bennett M, Deakos M, Stevens G, McGregor F, et al. Mobula alfredi (amended version of 2011 assessment). The IUCN Red List of Threatened Species. 2018; e.T195499A126665723

25. Marshall A, Barreto R, Carlson J, Fernando D, Fordham S, Francis MP, et al. Mobula birostris. The IUCN Red List of Threatened Species. 2020; e.T198921A68632946

26. Croll DA, Dewar H, Dulvy NK, Fernando D, Francis MP, Galván-Magaña F, et al. Vulnerabilities and fisheries impacts: the uncertain future of manta and devil rays. Aquat Conserv Mar Freshw Ecosyst. 2016; 26. https://doi.org/10.1002/aqc.2591

27. Stewart JD, Jaine FRA, Armstrong AJ, Armstrong AO, Bennett MB, Burgess KB, et al. Research Priorities to Support Effective Manta and Devil Ray Conservation. Front Mar Sci. 2018; 5. https://doi.org/10.3389/fmars.2018.00314
33. Anderson RC, Adam MS, Goes JI. From monsoons to mantas: Seasonal distribution of Manta alfredi in the Maldives. Fish Oceanogr. 2011; 20: 104–113. https://doi.org/10.1111/j.1365-2419.2011.00571.x

34. Harris JL, McGregor PK, Oates Y, Stevens GMW. Gone with the wind: Seasonal distribution and habitat use by the reef manta ray (Mobula alfredi) in the Maldives, implications for conservation. Aquat Conserv Mar Freshw Ecosyst. 2020; 30: 1649–1664. https://doi.org/10.1002/aqc.3350

35. Maldivian Manta Ray Project (MMRP). Oceanic Manta Ray Summary report 2019. The Manta Trust 2019. [cited 2020 November 18]. Available from: https://www.mantatrust.org/resources

36. Kashiwagi T, Marshall AD, Bennett MB, Ovenden JR. Habitat segregation and mosaic sympatry of the two species of manta ray in the Indian and Pacific Oceans: Manta alfredi and M. birostris. Mar Biodivers Rec. 2011; 4: 1–8 https://doi.org/10.1093/mbio/200109047

37. Stewart JD, Beale CS, Fernando D, Sianipar AB, Burton RS, Semmens BX, et al. Spatial ecology and conservation of Manta birostris in the Indo-Pacific. Biol Conserv. 2016; 200: 178–183. https://doi.org/10.1016/j.biocon.2016.05.016

38. Cressey RF, Lachner EA. The Parasitic Copepod Diet and Life History of Diskfishes (Echeneidae). Source: Copeia. 1970; 310–318.

39. O'Toole B. Phylogeny of the species of the superfamily Echeneoidea (Periformes: Carangoidae: Echeneidae, Rachycentridae, and Coryphaenidae), with an interpretation of echeneid hitchhiking behaviour. Can J Zool. 2002; 80: 596–623. https://doi.org/10.1139/z02-031

40. Kenaley CP, Stote A, Ludt WB, Chakrabarty P. Comparative Functional and Phylogenomic Analyses of Host Association in the Remoras (Echeneidae), a Family of Hitchhiking Fishes. Integr Org Biol. 2019; 1. https://doi.org/10.1093/IOB/doz007 PMID: 33793688

41. Flammang BE, Marras S, Anderson EJ, Lehmkohl O, Mukherjee A, Cade DE, et al. Remoras pick where they stick on blue whales. J Exp Biol. 2020; 223. https://doi.org/10.1242/jeb.226654 PMID: 33115921

42. Solleliet-Ferreira S, Macena BCL, Laglbauer BJL, Sobral AF, Afonso P, Fontes J.icklefin devilray and common remora prey jointly on baitfish. Environ Biol Fishes. 2020; 103: 993–1000. https://doi.org/10.1007/s10641-020-00990-9

43. Kordas RL, Harley CDG, O'Connor MI. Community ecology in a warming world: The influence of temperature on interspecific interactions in marine systems. J Exp Mar Bio Ecol. 2011; 400: 218–226. https://doi.org/10.1016/j.jembe.2011.02.029

44. Kronfeld-Schor N, Visser ME, Salis L, van Gils JA. Chronobiology of interspecific interactions in a changing world. Philos Trans R Soc B Biol Sci. 2017; 372(1734). https://doi.org/10.1098/rstb.2016.0248 PMID: 28993492

45. Stevens G, Froman N. The Maldives Archipelago. 2nd ed. In: Sheppard C, editor. World seas: an environmental evaluation: volume II: the Indian Ocean to the Pacific. 2nd ed. London: Academic Press; 2018. pp. 211–236.

46. Marshall AD, Compagno L, Bennett M. Redescription of the genus Manta with resurrection of Manta alfredi (Krefft, 1868) (Chondrichthyes; Myliobatoidei; Mobulidae). Zootaxa. 2009; 2301(1): 1–28. https://doi.org/10.11646/zootaxa.2301.1.1

47. Stevens G, Hawkins JP, Roberts CM. Courtship and mating behaviour of manta rays Mobula alfredi and M. birostris in the Maldives. J Fish Biol. 2018; 93: 344–359. https://doi.org/10.1111/jfb.13768 PMID: 3006396

48. Peacock LR, Stevens GMW, Daly R, Keating Daly CA, Collin SP, Nogue's J, et al. Regional Movements of Reef Manta Rays (Mobula alfredi) in Seychelles Waters. Front Mar Sci. 2020; 7: 1–17. https://doi.org/10.3389/fmars.2020.00548 PMID: 32802822

49. Froese R, Pauly D. FishBase. World Wide Web electronic publication. 2020 [cited 2021 March 28]. Available from: http://www.fishbase.org.

50. Chemos KL, Gutter AS, Blomberg SP, Marshall NJ. Blue and Yellow Signal Cleaning Behavior in Coral Reef Fishes. Curr Biol. 2009; 19: 1283–1287. https://doi.org/10.1016/j.cub.2009.06.028 PMID: 19592250

51. R Development Core Team. R: The R Project for Statistical Computing. In: Vienna, Austria: R Foundation Statistical Computing [Internet]. 2018 [cited 2 Dec 2020]. Available: https://www.r-project.org/

52. Bates AD, Maechler M, Bolker B, Walker S, Haubro R, Christensen B, et al. Linear Mixed-Effects Models using “Eigen” and S4 Package ’mem’. 2020.

53. Millar RB, Anderson MJ. Remedies for pseudoreplication. Fish Res. 2004; 70: 397–407. https://doi.org/10.1016/j.fishres.2004.08.016
55. Burnham K., Anderson D. Model Selection and Multimodel Inference: A Practical Information-Theoretic Approach (2nd ed.). Ecological Modelling. New York: Springer; 2002. https://doi.org/10.1016/j.ecolmodel.2003.11.004

56. Kamil Bartoń. Multi-Model Inference Package ‘MuMn.’ 2018.

57. Florian A. Package ‘DHARMa’ Residual Diagnostics for Hierarchical (Multi-Level / Mixed) Regression Models. 2021.

58. Bachman B. Life History Ecology of the Sharksucker *Echeneis naucrates*, in the Gulf of Mexico. PhD Thesis, The George Mason University. 2015. Available from: http://mars.gmu.edu/handle/1920/9801.

59. Brunnschweiler JM, Vignaud TM, Côté IM, Majkovič A. The costs of cohabiting: the case of sharksuckers (*Echeneis naucrates*) and their hosts at shark provisioning sites. Ecology. 2020; 101: 1–4. https://doi.org/10.1002/ecy.3160 PMID: 32799316

60. Braun CD, Skomal GB, Thorrold SR, Berumen ML. Diving Behavior of the Reef Manta Ray Links Coral Reefs with Adjacent Deep Pelagic Habitats. PLoS One. 2014; 9. https://doi.org/10.1371/journal.pone.0088170 PMID: 24516605

61. Jaine FRA, Rohner CA, Weeks SJ, Couturier LIE, Bennett MB, Townsend KA, et al. Movements and habitat use of reef manta rays off eastern Australia: Offshore excursions, deep diving and eddy affinity revealed by satellite telemetry. Mar Ecol Prog Ser. 2014. https://doi.org/10.3354/meps10820 PMID: 27239082

62. Stewart JD, Smith TR, Marshall G, Abernathy K, Fonseca-Ponce IA, Froman N, et al. Novel applications of animal-borne Crittercams reveal thermocline feeding in two species of manta ray. Mar Ecol Prog Ser. 2019; 632: 145–158. https://doi.org/10.3354/meps13148

63. Lassauce H, Chateau O, Erdmann M V., Wantiez L. Diving behavior of the reef manta ray (*Mobula alfredi*) in New Caledonia: More frequent and deeper night-time diving to 672 meters. PLoS One. 2020; 15(3). https://doi.org/10.1371/journal.pone.0228815 PMID: 32187197

64. Brunnschweiler JM, Baensch H, Pierce SJ, Sims DW. Deep-diving behaviour of a whale shark *Rhincodon typus* during long-distance movement in the western Indian Ocean. J Fish Biol. 2009; 74: 706–714. https://doi.org/10.1111/j.1095-8649.2008.02155.x PMID: 20735591

65. Tyminski JP, De La Parra-Venegas R, Cano JG, Hueter RE. Vertical movements and patterns in diving behavior of whale sharks as revealed by pop-up satellite tags in the Eastern Gulf of Mexico. PLoS One. 2015; 10. https://doi.org/10.1371/journal.pone.0142156 PMID: 26580405

66. Germanov ES, Bejder L, Chabanee DB, Dharmadi D, Hendrawan IG, Marshall AD, et al. Contrasting habitat use and population dynamics of reef manta rays within the Nusa Penida Marine Protected Area, Indonesia. Front Mar Sci. 2019; 6: 215. https://doi.org/10.3389/fmars.2019.00215

67. Harris JL, Hosegood P, E R, Embling C, Hilbourne S, Stevens G. The tide waits for no manta: Fine-scale oceanographic drivers of reef manta ray, *Mobula alfredi*, visitation patterns at a feeding aggregation site. Ecol Evol. 2021. (In Review).

68. Drapella Z. The hitchhiker’s guide to the Maldives: The echeneid-host association between remoras and manta mays. M.Sc Thesis, The University of Southampton. 2016. Available: https://www.mantatrust.org/resources

69. Wallman HL, Bennett WA. Effects of parturition and feeding on thermal preference of Atlantic stingray, *Dasyatis sabina* (Lesueur). Environ Biol Fishes. 2006; 259–267. https://doi.org/10.1007/s10641-006-0025-1

70. Hight B V., Lowe CG. Elevated body temperatures of adult female leopard sharks, *Triakis semifasciata*, while aggregating in shallow nearshore embayments: Evidence for behavioral thermoregulation? J Exp Mar Bio Ecol. 2007; 352: 114–128. https://doi.org/10.1016/j.jembe.2007.07.021

71. Bernal D, Lowe CG. Field studies of elasmobranch physiology. 1st ed. In: Shadwick R, Farrell A, Braum C, editors. Physiology of elasmobranch fishes: structure and interaction with environment. Amsterdam: Elsevier Ltd; 2016.

72. Nakajima H, Kawahara H, Takamatsu S. The breeding behavior and the behavior of larvae and juveniles of the sharksucker, *Echeneis naucrates*. Japanese J Ichthyol. 1987; 34(1): 66–70. https://doi.org/10.1007/BF02904145

73. Winemiller KO. Life history strategies, population regulation, and implications for fisheries management. Can J Fish Aquat Sci. 2005; 62: 872–885. https://doi.org/10.1139/f05-040

74. Alexander RL. Evidence of brain-warming in the mobulid rays, *Mobula tarapacana* and *Manta birostris* (Chondrichthyans: Elasmobranchii: Batoidea: Myliobatiformes). Zool J Linn Soc. 1996; 118(2): 151–164. https://doi.org/10.1007/BF02904145

75. Thums M, Meekan M, Stevens J, Wilson S, Polovina J. Evidence for behavioural thermoregulation by the world’s largest fish. J R Soc Interface. 2013; 10(78). https://doi.org/10.1098/rsif.2012.0477 PMID: 23075547
76. Croll D, Newton K, Weng K, Galván-Magaña F, O’Sullivan J, Dewar H. Movement and habitat use by the spine-tail devil ray in the Eastern Pacific Ocean. Mar Ecol Prog Ser. 2012; 465: 193–200. https://doi.org/10.3354/meps09900

77. Hino H, Kitagawa T, Matsumoto T, Aoki Y, Kimura S. Changes to vertical thermoregulatory movements of juvenile bigeye tuna (Thunnus obesus) in the northwestern Pacific Ocean with time of day, seasonal ocean vertical thermal structure, and body size. Fish Oceanogr. 2019; 28: 1–13. https://doi.org/10.1111/fo.12417

78. Sancho G. Predatory behaviors of Caranx melampygus (Carangidae) feeding on spawning reef fishes: A novel ambushing strategy. Bull Mar Sci. 2000; 66(2): 487–496.

79. Lédee EJ, Heupel MR, Tobin AJ, Simpfendorfer CA. Movements and space use of giant trevally in coral reef habitats and the importance of environmental drivers. Anim Biotelemetry. 2015; 3(1): 6. https://doi.org/10.1186/s40317-015-0024-0

80. Garcia Ramírez CB, Posada C. First approach to the trophic ecology and diet of the rainbow runner, Elagatis bipinnulata (Quoy & Gaimard) (Pisces: Carangidae), in the central Columbian Caribbean. Acta Biológica Colomb. 2014; 19: 309. https://doi.org/10.15446/abc.v19n2.41294

81. Wells RJ, Cowan JH, Fry B. Feeding ecology of red snapper Lutjanus campechanus in the northern Gulf of Mexico. Mar Ecol Prog Ser. 2008; 361: 213–225. https://doi.org/10.3354/meps07425

82. Aronson R. Foraging behavior of the west Atlantic trumpetfish, Aulostomus maculatus: use of large, herbivorous reef fishes as camouflage. Bull Mar Sci. 1983;33(1).

83. Grandcourt EM, Al Abdessalam TZ, Francis F, Al Shamsi A. Population biology and assessment of representatives of the family Carangidae: Carangoides bajad and Gnathanodon speciosus (Forsskål, 1775), in the Southern Arabian Gulf. Fish Res. 2004; 69: 331–341. https://doi.org/10.1016/j.fishres.2004.06.008

84. Kuitert R. Fishes of the Maldives: Indian Ocean. Cairns, Australia: Atoll Editions; 2019.

85. Moreno B, Gonzalez-Pestana A. Southernmost record of the giant manta ray mobula birostris (Walbaum, 1792) in the eastern pacific. Mar Biodivers Rec. 2017; 10(1): 27. https://doi.org/10.1186/s41200-017-0130-1

86. Harley CDG, Randall Hughes A, Hultgren KM, Miner BG, Sorte CJB, Thorber CS, et al. The impacts of climate change in coastal marine systems. Ecol Lett. 2006; 9(2): 228–241. https://doi.org/10.1111/j.1461-0248.2005.00871.x PMID: 16958887

87. Joseph P V., Simon A. Weakening trend of the southwest monsoon current through peninsular India from 1950 to the present. Curr Sci. 2005; 89(4): 687–694.

88. Roxy MK, Modi A, Murtugudde R, Valsala V, Panickal S, Prasanna Kumar S, et al. A reduction in marine primary productivity driven by rapid warming over the tropical Indian Ocean. Geophys Res Lett. 2016; 43: 826–833. https://doi.org/10.1002/2015GL066979

89. Dhunya A, Huang Q, Aslam A. Coastal Habitats of Maldives: Status, Trends, Threats, and potential conservation Strategies. Int J Environ Eng Res. 2017; 8: 47–62.

90. Dunkley K, Ward AJW, Perkins SE, Cable J. To clean or not to clean: Cleaning mutualism breakdown in a tidal environment. Ecol Evol. 2020; 10: 3043–3054. https://doi.org/10.1002/ece3.6120 PMID: 3221175

91. Montoya JM, Raffaelli D. Climate change, biotic interactions and ecosystem services. Philos Trans R Soc B Biol Sci. 2010; 365. https://doi.org/10.1098/rstb.2010.0114 PMID: 20513709

92. Halpern BS, Frazier M, Afflerbach J, Lowndes JS, Micheli F, O’Hara C, et al. Recent pace of change in human impact on the world’s ocean. Sci Rep. 2019; 9. https://doi.org/10.1038/s41598-019-47201-9 PMID: 31406130