Effects of Nitrification Inhibitors on Nitrogen Dynamics and Ammonia Oxidizers in Three Black Agricultural Soils

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Abstract: The application of nitrification inhibitors (NIs) based on ammonium (NH$_4^+$) is considered to be an efficient way to reduce nitrogen (N) loss by delaying the nitrification process through influencing ammonia-oxidizing archaea (AOA) and ammonia-oxidizing bacteria (AOB). However, the inhibition mechanisms of NIs on AOA and AOB in different soils remain unclear. Hence, we explored the two commonly used NIs (3, 4-dimethylepyrazole phosphate (DMPP) and dicyandiamide (DCD) and their combination (DMPP + DCD) on the soil nitrification and abundance of ammonia oxidizers (AOA and AOB) in three black soils (HLJ, NA, and DA) with different physicochemical properties using a 120-day incubation experiment. The results demonstrated that NIs significantly increased NH$_4^+$-N concentrations and decreased NO$_3^-$-N concentrations in all three tested soils. There was no significant difference in inhibiting nitrification in HLJ among all NI treatments, while DCD was more efficient in NA, DMPP + DCD had better efficiency in DA. The potential nitrification rate (PNR) was greatly decreased by NIs addition, and PNR was significantly positively correlated with AOB (p < 0.05). AOA was dominant in the acid soil. All NI treatments significantly inhibited soil nitrification through inhibiting the growth of AOB in the two soils with higher pH. The abundance of AOA and AOB was significantly correlated with different soil types (positively correlated with soil pH, and negatively correlated with organic matter). Moreover, soil pH and soil organic matter were considered to be the most important factors influencing the inhibition efficiency of NIs and the abundance of AOA and AOB. The application of the NIs combination (DMPP + DCD) was considered to be the most cost-effective way to inhibit soil nitrification in soil with higher pH and lower SOM, which provides a theoretical basis for a field experiment.

Keywords: nitrification; ammonia oxidizing archaea (AOA); ammonia oxidizing bacterial (AOB); pH; soil organic matter

1. Introduction

It is well known that nitrogen (N) is one of the most important nutrients for crops, and applying N-rich fertilization is a widespread practice in agriculture to ensure the higher yield and quality of plant. Soil nitrification is a critical and integral part of the N cycle, during which a relatively immobile ammonium N (NH$_4^+$) form is converted into highly mobile nitrate N (NO$_3^-$) form [1]. However, excessive application of ammonium-based fertilizer results in ammonia volatilization [2], NO$_3^-$ leaching and N$_2$O emission, which bring a series of problems such as the lower crop yield, the environmental pollution (e.g., acidification and eutrophication), and climate change (global warming) [3,4]. The application of nitrification inhibitors (NIs) with ammonium-based fertilizer is considered to be one of the most efficient ways to reduce N loss and mitigate environmental pollution by reducing nitrification.
NIs are chemical compounds that delay soil nitrification and decrease the emissions of the potent greenhouse gas nitrous oxide (N$_2$O), leaching, and denitrification [5]. Nitrification inhibitors such as 3, 4-dimethylepyrazole phosphate (DMPP) and dicyandiamide (DCD) are the most widely used in agricultural soils [6]. A field experiment demonstrated that DCD with urea significantly reduced NO$_3^-$ leaching by 58.5% and 36.2% in soils with pH 5.12 and 7.71, respectively [7]. Another field study showed that DCD with urea increased the nitrogen use efficiency by 10% [8]. A pot experiment indicated that DMPP and DCD combined with urea kept the higher NH$_4^+$-N content longer in brown soil than that in cinnamon soil [9]. Obviously, the previous studies mostly focused on the effect of single NIs combined with urea, but lacked comprehensive research on the effect of NIs combination with ammonium-based nitrogen fertilizer. Additionally, different NIs have different properties and price [6], it is essential to study the effect of combination NIs to find a more cost-effective way to inhibit soil nitrification.

The first and rate-limiting step of nitrification is the ammonia oxidation (from NH$_4^+$ to NH$_2$OH), which is catalyzed by the key enzyme ammonia monoxygenase (AMO). AMO is encoded by the amoA gene and performed by ammonia-oxidizing archaea (AOA) or ammonia-oxidizing bacteria (AOB) [10,11]. The core mechanism of nitrification has been reported that DMPP and DCD are deemed to the copper chelating of AMO [12]. DMPP binds indiscriminately to the complex of membrane-bound proteins that include AMO and that catalyze the reaction of the enzyme AMO [13], while DCD inhibits the activity of the enzyme AMO by interfering with the electron transport in the cytochromes of AMO [14]. However, there is little research to study the inhibitory mechanism of NI combinations in order to better utilize NIs.

Previously, many experiments have explored the influence of NIs on the abundance and community structure of AOA and AOB [6,9,15]. However, the mechanism of NIs is still unclear because of the variable responses of AOA and AOB to NIs’ application. This could be due to the different metabolic pathways and the different ecological niches between AOA and AOB to NIs [16]. Soil pH is one of the key factors to drive the niche partitioning of AOB and AOA [6]. AOB grew preferentially in soils with relatively high pH and high NH$_3$ concentration, while AOA grew preferentially in acid soil with lower nutrient availability [17]. An incubation experiment suggested that NIs inhibited soil nitrification mainly by impairing the AOB amoA gene rather than the AOA amoA gene [18]. However, another experiment showed that NIs decreased both the abundance of AOB and AOA amoA gene [9]. However, in an acid soil where AOA was more active than AOB, NIs only inhibited AOA [19]. Soil pH, which not only affects the inhibitor effect indirectly by affecting ammonia-oxidizing microorganisms, but also directly affects the inhibitor performance. For example, higher absorption of DCD has been found in a more alkaline soil [20], and significant higher efficacy of NIs was demonstrated in an acid grassland soil [21]. Due to the various efficiency of NIs, it is urgent to demonstrate the effect of NIs on ammonia oxidizers in different black soils with different physicochemical properties.

Black soils are classified as Mollisols in the United States System of soil taxonomy, which are the major agricultural soils in northeast China [22]. We hypothesized that (i) soils with different properties would result different NIs and their combination would have different effects on the nitrification process and that (ii) AOA and AOB resident in the three soils would respond differently to different NIs and their combination. Therefore, this study will be conducted on AOA and AOB abundance on three black soils with different physicochemical properties to determine the effect of NIs (DMPP, DCD, and their combination DMPP + DCD) on soil inorganic nitrogen concentrations, potential nitrification rate (PNR), and the abundance of AOA and AOB. This study will help us better design the application of NIs and provide a theoretical basis for a field experiment.
2. Material and Methods

2.1. Soil Samples

Three soils were used for the incubation experiment: an acid soil at 853 farm (HLJ: 46°32’ N, 132°15’ E), Heilongjiang Province of China; and a neutral soil and an alkaline soil at Nong’an (NA: 44°43’ N, 125°18’ E) and Da’an (DA: 45°31’ N, 123°56’ E), respectively, Jilin Province of China. Three sites are typical agricultural soils in northeast China. These soils collected from different sites had different physico-chemical properties (Table 1).

Table 1. Physico-chemical properties of three different soils.

| Soil Property | HLJ     | NA      | DA      |
|---------------|---------|---------|---------|
| pH            | 5.44 ± 0.13 | 7.66 ± 0.07 | 9.94 ± 0.17 |
| Total C (g kg\(^{-1}\)) | 30.31 ± 1.11 | 18.94 ± 0.91 | 17.47 ± 0.32 |
| Total N (g kg\(^{-1}\)) | 2.63 ± 0.03 | 1.66 ± 0.07 | 0.92 ± 0.15 |
| NH\(_4^+\)-N (mg kg\(^{-1}\)) | 18.69 ± 1.05 | 27.83 ± 3.46 | 44.44 ± 3.48 |
| NO\(_3^-\)-N (mg kg\(^{-1}\)) | 80.68 ± 1.46 | 132.73 ± 2.19 | 24.33 ± 2.16 |
| Available P (mg kg\(^{-1}\)) | 48.40 ± 2.13 | 18.42 ± 0.56 | 15.43 ± 0.32 |
| Available K (mg kg\(^{-1}\)) | 401.45 ± 34.27 | 344.04 ± 19.23 | 375.28 ± 24.33 |
| SOM (g kg\(^{-1}\)) | 52.25 ± 1.91 | 32.65 ± 1.57 | 30.12 ± 0.54 |
| Clay%          | 12.3     | 37.3    | 60.6    |
| Silt%          | 44.3     | 52.2    | 37.3    |
| Sand%          | 43.4     | 10.4    | 2.1     |
| Texture class  | loam    | silt clay | clay   |

Total C: total carbon; Total N: total nitrogen; NH\(_4^+\)-N: ammonium nitrogen; NO\(_3^-\)-N: nitrate nitrogen; P: phosphorus; K: potassium; SOM: soil organic matter.

For all sites, soil samples were collected from 0 to 20 cm depth, then thoroughly mixed, and immediately transported to the laboratory. The soil samples were passed through a 2 mm sieve to remove coarse plant debris and stones, and stored at room temperature before using.

2.2. Soil Incubation Experiment

Five treatments were set during incubation experiment: (1): CK: no fertilizer and NIs; (2) N: ammonium sulfate; (3) DMPP: ammonium sulfate + 3, 4-dimethylepyrazole phosphate; (4) DCD: ammonium sulfate + dicyandiamide; (5) DMPP + DCD: ammonium sulfate + 3, 4-dimethylepyrazole phosphate + dicyandiamide. Each treatment had three replicates. N fertilizer was applied at 0.5 g N kg\(^{-1}\) dry soil. The application rate of DMPP and DCD were 0.5% and 4%, respectively, on the w/w basis of N, and the application rate of every single NI was reduced by 50% in NIs combination treatment. All soils were pre-incubated at 25 °C for one week at 20% water holding capacity (WHC). Then, 1 kg (dry soil) was placed into a column (17 cm in diameter and 11.5 cm in height). All columns were incubated in the dark at 25 ± 1 °C for 120 days. The soil moisture was adjusted to 60% WHC with deionized water. During the incubation period, deionized water was regularly added to maintain soil water contents equivalent to 60% WHC.

2.3. Soil Sampling and Analysis

During the incubation period, soil samples were taken from each treatment at specific intervals (1, 3, 5, 7, 14, 21, 28, 35, 45, 70, 85, 100, 110, and 120 days). The soil samples were divided into two parts, one part was stored at 4 °C for the determination of soil inorganic nitrogen and potential nitrification rate (PNR) (1, 7, 28, 70, and 100 days); another was frozen at −80 °C for molecular analysis (1, 7, 28, 70, and 100 days).

Soil pH was determined using a pH meter (S200, Mettler Toledo, Switzerland) with a ratio of 1:2.5 (soil:water). Total C and total N of soil were determined by dry combustion using an Elemental Analyzer (Vario EL III, Hanau, Germany) [23]. The soil organic matter content was measured by the potassium dichromate (K\(_2\)Cr\(_2\)O\(_7\)) volumetric method [24]. Soil available phosphorus (AP) was determined by the molybdenum blue method on sodium bicarbonate extracts, and soil available potassium (AK) was determined by extraction with
ammonium acetate. Soil moisture content was determined by oven-drying at 105 °C for 8 h. Soil inorganic nitrogen (NH$_4^+$-N and NO$_3^-$-N) was extracted with 2 mol L$^{-1}$ KCl and determined on a continuous flow analyzer (AA III, Norderstedt, Germany) [23]. The soil PNR was determined by using the chlorate inhibition method [25].

2.4. Soil DNA Extraction and Quantitative PCR (qPCR) Analysis

Soil DNA was extracted from 0.3 g (wet weight) of soil using the Power Soil DNA Isolation Kit (for soil) (MoBio Laboratories Inc., San Diego, CA, USA) following the manufacturer’s instructions. The DNA concentration was measured using a NanoDrop-2000 Spectrophotometer (Thermo Fisher Scientific, Waltham, MA, USA) and the quality checked using 1% agarose gel electrophoresis. All extracted DNA was diluted with sterilized MilliQ water (1:10) to reduce potential PCR inhibition, and stored at −20 °C before amplification.

The abundances of bacterial and archaea ammonia monoxygenase (amoA) were quantified by real-time quantitative PCR (qPCR) on an ABI 7500 thermocycler (Applied Biosystems, Waltham, MA, USA). The specific primer combinations and thermal cycling conditions are listed in Table 2. Each PCR was performed in a 20 µL reaction mixture consisting of 1 µL of each primer, 10 µL SYBR® Premix Ex Taq (TaKaRa, Tokyo, Japan), 2 µL DNA template, and the residual volume was replenished with deionized water. Standard curves were generated using 10-fold serial dilutions of plasmids containing correct inserts of the target genes. Melting curve analysis was performed between 72 and 95 °C at the end of each amplification assay to evaluate the specificity of qPCR products. Real-time PCR was performed in triplicate and amplification efficiencies of 92.2–105.3% were obtained, with the correlation coefficient (R$^2$) of the determination ranged from 0.95 to 0.99.

Table 2. Primer sets and PCR profiles used in the real-time PCR.

| Target Group | Primer Set      | Sequence (5′-3′)             | Annealing Temperature (°C) | Reference |
|--------------|----------------|------------------------------|---------------------------|-----------|
| AOA          | Arch-amoAF     | STAATGGTCTGGCTTAGACG         | 58 °C                     | [26]      |
|              | Arch-amoAR     | GCGGCCATCCATCTGTATGT         |                           |           |
|              | AmoA-1F        | GGGGTTTTCTACTGGTG            | 60 °C                     | [27]      |
|              | AmoA-2R        | CCCCTCKGSAAAGGCTTTCTTC      |                           |           |

2.5. Statistical Analysis

Data were presented as the means of the results of the three replicates. One way analysis of variance (ANOVA) was conducted to test the effects of all treatments at each incubation time. Significant differences between the means were analyzed using Tukey’s multiple comparisons at the 0.05 probability level. The effect of different treatments, soils with different pH level, and their interaction on the abundances of the functional genes, soil NH$_4^+$, NO$_3^-$, and PNR was examined by two-way repeated ANOVA. In addition, the correlations among the abundances of the functional genes, soil NH$_4^+$ and NO$_3^-$ concentration, PNR, NIs, and soils with different pH level were analyzed by using the Pearson correlation test with Origin 2021. All statistical analyses were conducted by using the statistical software SPSS 22.0.

3. Results

3.1. Soil Inorganic Nitrogen Concentrations in Three Soils

The dynamic changes in NH$_4^+$-N concentrations varied in three black soils with different during the incubation time. During 120 days of incubation, NH$_4^+$-N content gradually decreased (Figure 1A,C,E). The NH$_4^+$-N concentration in N treatment rapidly decreased by 15–30% after day 3, though still 92–97% higher than that in the control only in HLJ, and no significant difference between CK and N treatments after day 14 and 70 in NA and DA, respectively (Figure 1A,C,E; Table S1). In HLJ, no significant difference was found in NI treatments among the entire incubation time except for day 3 and maintained NH$_4^+$-N concentration > 346.35 mg kg$^{-1}$ at least 120 days (Table S1). There was no significant
difference in NH$_4^+$-N concentration in NIs treatments before day 14 and 35 in NA and DA, respectively (Figure 1C,E, p > 0.05). In NA, there was no significant difference in NH$_4^+$-N concentration in NI treatments before day 14. DCD remained NH$_4^+$-N concentration > 55.13 mg kg$^{-1}$ for 100 days (Table S1), which was longer than DMPP and DMPP+DCD (Figure 1C). All NI treatments maintained NH$_4^+$-N concentration > 255.98 mg kg$^{-1}$ for at least 120 days in DA, especially DMPP + DCD, which exhibited more efficiency in inhibiting nitrification (NH$_4^+$-N > 416.20 mg kg$^{-1}$) (Table S1). The results of two-way ANOVA indicated that both NIs and soils with different pH level had a significant impact on NH$_4^+$-N concentrations (Table 3).

![Dynamic changes in ammonium nitrogen (NH$_4^+$-N, left) and nitrate nitrogen (NO$_3^-$-N, right) in HLJ (A,B), NA (C,D), and DA (E,F) among different treatments during 120 days incubation. Error bars represent standard deviations of means (n = 3). HLJ: soil with pH 5.44; NA: soil with pH 7.66; DA: soil with pH 9.96. Treatments: CK: no fertilizer and NIs; N: ammonium sulfate; DMPP: ammonium sulfate + 3, 4-dimethylepyrazole phosphate; DCD: ammonium sulfate + dicyandiamide + 3, 4-dimethylpyrazole phosphate; DMPP + DCD: ammonium sulfate + 3, 4-dimethylpyrazole phosphate + dicyandiamide. The same below.

**Figure 1.** Dynamic changes in ammonium nitrogen (NH$_4^+$-N, left) and nitrate nitrogen (NO$_3^-$-N, right) in HLJ (A,B), NA (C,D), and DA (E,F) among different treatments during 120 days incubation. Error bars represent standard deviations of means (n = 3). HLJ: soil with pH 5.44; NA: soil with pH 7.66; DA: soil with pH 9.96. Treatments: CK: no fertilizer and NIs; N: ammonium sulfate; DMPP: ammonium sulfate + 3, 4-dimethylepyrazole phosphate; DCD: ammonium sulfate + dicyandiamide + 3, 4-dimethylpyrazole phosphate; DMPP + DCD: ammonium sulfate + 3, 4-dimethylpyrazole phosphate + dicyandiamide. The same below.

**Table 3.** Two-way ANOVA (p < 0.05) examining the effect of nitrification inhibitors (NIs), soils with different pH level (S) and their interaction (NIs × S) on AOA abundance, AOB abundance, NH$_4^+$-N, NO$_3^-$-N, and potential nitrification rate (PNR) during the incubation.

| Factors     | DF | AOA | AOB | NH$_4^+$-N | NO$_3^-$-N | PNR |
|-------------|----|-----|-----|------------|------------|-----|
| NIs         | 3  | 1.3 | 5.9 | 7.4        | 1.0        | 12.7|
| S           | 2  | 273.4 | 88.2 | 41.9 | 60.0           | 45.3|
| NIs × S     | 6  | 1.9 | 4.0 | 1.3        | 4.3        | 7.1  |

DF: degree of freedom of two independents samples; F value: the ratio of mean squares of two independents samples; p value: the index of differences between the control group and the experimental group; n.s.: not significant; ** indicates significance at p < 0.01; *** indicates significance at p < 0.001.

The NO$_3^-$-N concentrations in three soils gradually increased during the incubation time (Figure 1B,D,F). N treatments had higher concentrations of NO$_3^-$-N than those in the CK treatments in three soils, which was consistent with the results of NH$_4^+$-N concentration. In HLJ and DA, the NO$_3^-$-N concentrations in the three NI treatments were significantly lower than those in the N treatments during incubation time (Figure 1B,F, p < 0.05), while in DA, the NO$_3^-$-N concentrations in the NI treatments except for DCD remained relatively stable during the incubation (Figure 1F). In NA, three NIs treatments
in NO$_3^-$-N concentrations were lower than those in N treatment before 70 days, while the NO$_3^-$-N concentrations in N treatment was equal or lower than those in NI treatments except for DCD (Figure 1D, $p < 0.05$).

3.2. Soil Potential Nitrification Rate (PNR)

The results of two-way ANOVA indicated that the NIs, soils with different pH levels, and their interaction all had a significant influence on PNR (Table 3, $p < 0.05$). In HLJ and DA, there was no significant difference was found in PNR between N and the NI treatments at day 1 (Figure 2A, $p < 0.05$), while higher PNR in N treatment was observed in NA and DA at day 28 than those in the NI treatments ($p < 0.05$). During the incubation time, the PNR in the N treatment was equal to or lower than those in the NI treatments in HLJ, expect DMPP + DCD at day 100. While in NA and DA, the PNR in N treatment was higher or equal to that in the NI treatments, except DMPP + DCD at day 100 in NA (Figure 2B,C, $p < 0.05$). For each NI treatment, all NI treatments significantly decreased the PNR for each soil, especially in DA (Figure 2, $p < 0.05$).

![Figure 2. Potential nitrification rate (PNR) in the HLJ (A), NA (B), and DA (C) under different treatments on each sampling day during the incubation. The different letters above the figures of the same sampling day indicate significant differences between treatments at $p < 0.05$ by Tukey test. The same below.](image)

3.3. Abundance of the AOA and AOB amoA Gene

The abundance of AOA and AOB among different treatments in the three tested soils during the incubation period was shown in Figure 3. The results of two-way ANOVA suggested that AOA abundance could be significantly affected by soils with a different pH level, but NIs, soils with different pH level, and their interaction all could significantly influence AOB abundance (Table 3). In HLJ, N treatment significantly decreased AOA abundance, while increased AOB abundance compared with CK (Figure 3A, $p < 0.05$). NI treatments significantly increased AOA abundance compared with N treatment, but there was no significant impact in AOB abundance between NIs and N treatments (Figure 3B, $p < 0.05$). In NA, a significant increase in both AOA and AOB abundance in N treatment compared with CK (Figure 3C, $p < 0.05$), NI treatments significantly decreased AOA and AOB abundance compared with N treatment (Figure 3C,D, $p < 0.05$). There was no significant increase in AOA abundance in NI treatment than that in N treatment (Figure 3C, $p < 0.05$), but NIs significantly decreased AOB abundance (Figure 3D, $p < 0.05$). In DA, no significant difference in AOA abundance was found between N and CK treatments during
the entire incubation time (Figure 3E, \( p < 0.05 \)). NI treatments increased AOA abundance significantly only at day 7 compared with AS, while no significant increase was found at other incubation times (Figure 3E, \( p < 0.05 \)). DMPP + DCD significantly increased AOB abundance, but no significant increase was found in the other two NIs treatments (DMPP, DCD) at day 1. During the latter period of the incubation, all NIs treatments significantly decreased AOB abundance, especially DMPP and DMPP + DCD (Figure 3F, \( p < 0.05 \)).

4. Discussion

In each soil, all N treatments had higher NH\(_4\)\(^+\)-N and NO\(_3\)\(^-\)-N contents than that in the treatments without N fertilizer (CK), indicating that ammonia sulfate application improved the content of soil inorganic nitrogen. In addition, the decrease in NH\(_4\)\(^+\)-N content and the increase in NO\(_3\)\(^-\)-N content in N treatments during the incubation time, showed that ammonia sulfate application promoted soil nitrification. However, the degree of soil nitrification varied in all three tested soils: the NH\(_4\)\(^+\)-N content kept higher during entire incubation period in HLJ (Figure 1A), while the NH\(_4\)\(^+\)-N content significantly decreased at day 14 and 70 in NA and DA, respectively, and slightly changed during the rest of the incubation period (Figure 1C,E). The findings are consistent with previous studies, which also showed nitrification in different soils evolved differently during incubation [17,28].

The main reason for this difference is that the tested soils had a different pH. Soil pH has been considered as a major factor regulating the nitrification process in soils, and that nitrification is rapid in soils with pH > 6.0 [2,29]. Soil nitrification was affected by the availability of the ammonium ion, which in turn was impacted by the quality of soil organic matter [29]. Additionally, the result of the Pearson correlation revealed that soil NH\(_4\)\(^+\)-N content had a negative correlation with soil pH (Figure 4), while it had a significantly positive correlation with SOM (Figure 4).
The possible explanation is that we had the same experiment environment (laboratory incubation), which had no nitrate leaching and crop uptake. The dynamic changes in all NI treatments on soil NH$_4^+$-N and NO$_3^-$-N concentrations during the incubation time (Figure 1). These results were substantiated by many researchers, who found DMPP and DCD could reduce soil nitrification [9,30,31]. Similarly, PNR also significantly decreased during NI treatments (Figure 2). These findings are consistent with the results of the Pearson-correlation analysis, which showed soil NH$_4^+$-N concentration had a significantly negative correlation with soil PNR, while soil NO$_3^-$-N concentration had a significantly positive correlation with soil PNR (Figure 4). Our experiment showed that NH$_4^+$-N concentrations maintained at higher levels in the NI treatments for at least 45 days ($p < 0.05$) in all three tested soils. The result is similar with the result of Zerulla et al. [32], who found the effect of DMPP on soil NH$_4^+$-N concentrations lasted for approximately 40 days. Dong et al. [33] reported that DMPP with urea was able to maintain lower NO$_3^-$-N for nearly 90 days, which is also consistent with our results (120 days). The possible explanation is that we had the same experiment environment (laboratory incubation), which had no nitrate leaching and crop uptake. The dynamic changes in all NI treatments on soil NH$_4^+$-N and NO$_3^-$-N concentrations were similar in HLJ and DA in the present study, which is consistent with the result of Yang et al. [34]. However, the reason for this phenomenon may be different. The lower nitrification happened in HLJ, whereas there was better efficiency in inhibiting soil nitrification in DA. Moreover, no significant differences were found in inorganic nitrogen among all NI treatments in HLJ, while DCD performed better in inhibiting nitrification than DMPP and DMPP + DCD in NA, and DMPP + DCD was more effective than DMPP and DCD in DA (Figure 1). A laboratory experiment also found that DCD was more effective than DMPP at soil with pH 7.6 [21]. The reason for this phenomenon may be attributed to the weak soil nitrification in HLJ and the low potential nitrification rate (PNR) in both NA and DA soils after adding NIs into the N fertilizer [11,16]. PNR has been used to demonstrate the ability of soil ammonium nitrogen oxidation to nitrate nitrogen [10]. The highest PNR with N treatment was in NA,

- $p<0.05$ ** $p<0.01$ *** $p<0.001$

Figure 4. Pearson correlation test among soil NH$_4^+$-N, NO$_3^-$-N, potential nitrification rate (PNR), functional genes (AOA, AOB) abundance, and soils with different soil properties (pH and SOM (soil organic matter)). The size of the circle and the shade of color represents the degree of relevance. The number is the correlation coefficient.
followed by DA, and the lowest PNR value was found in HLJ (Figure 2). This could be caused by the environment in NA, which was more favorable for ammonia oxidizers (AOA and AOB). In the NI treatments, the lowest PNR value was found in DA, indicating that NIs showed better efficiency in inhibiting ammonia oxidation, which is consistent with Feng et al. [35]. The higher efficiency of NIs in alkaline soil might be due to the fact that NIs were retained better and were more susceptible to nitrifier population in high pH soils than in low pH soils [36]. At the end of the incubation time, there was no significant difference in inorganic nitrogen concentrations among three NIs treatments in three soils. This phenomenon is mainly caused by decomposition of NIs [9] and the soil adsorption of them [37].

The effectiveness of NIs treatments (DMPP, DCD, and DMPP + DCD) in delaying soil nitrification differed among the three tested soils (Figure 1), which could be explained by the effect of the contrasting soil properties. One of these properties is soil pH, which has the potential to influence the mobility and degradation of the NIs in soils [20]. The results of Pearson correlation analysis indicated that soil pH had a significantly positive correlation with soil NO$_3^-$-N (Figure 4). The soils with higher pH increased the soil microbial activity and soil nitrification, which was more conducive to the role of NIs [16]. However, the higher microbial activity may result in the rapid biodegradation of NIs [38]. All NIs treatments maintained higher NH$_4^+$-N concentration in DA (pH: 9.94) than that in HLJ (pH: 5.44) and NA (pH: 7.66). Moreover, all NI treatments significantly delayed soil nitrification, but the different durations from different treatments were found in DA (Figure 1). Aside from soil pH, SOM was also considered to be one of the important factors to influence the inhibition efficiency [39]. The adsorption of SOM to NIs can reduce the loss of their volatilization, which is beneficial to maintain NIs in the soils. Meanwhile, NIs are protected by abiotic organisms, which reduce their biological activity to some extent [40]. DMPP was more efficient in inhibiting nitrification in NA than that in DA (Figure 1). Previous study has also revealed that the inhibitory effects of DMPP were much less in soils with higher SOM than in soils with lower SOM [41]. The possible explanation is that DMPP, as a type of heterocyclic compound, which can be adsorbed by SOM [42], thereby leads to low DMPP availability in soils [43].

The results of the Pearson correlation analysis showed that soil AOB abundance had an extremely significant positive correlation with soil PNR (Figure 4), indicating that AOB was more responsible for the ammonia oxidation [11]. However, the weak correlation between the soil AOA abundance and PNR could be due to the method for determination of PNR, in which we added excess NH$_4^+$ [1]. The results of the Pearson correlation test also revealed that soil pH and SOM were significantly correlated with AOA and AOB abundance (Figure 4). Soil pH was positively correlated with AOA abundance (Figure 4), indicating that the abundance of AOA decreased with decreasing pH values [44]. In our study, we found more AOA amoA genes in NA and DA than those in HLJ. The finding accords with the results of the Pearson correlation analysis, which indicated that AOA abundance had a significantly negative correlation with SOM (Figure 4). Most reports confirmed that AOA has been adopted to neutral or alkaline conditions [45,46]. In addition, the lower AOA and AOB abundance may contribute to the lower potential nitrification in HLJ, which is in line with recent studies [17]. Although AOA and AOB abundances were different among the three soils, N treatment significantly increased AOB abundance in HLJ and NA. Similar effects of N fertilizer on AOB abundance were reported under field experiment [47] and incubation conditions [48]. This could be due to AOB preferring high ammonia substrate conditions [49].

In HLJ, AOA abundance significantly decreased after the application of (NH$_4^+$)$_2$SO$_4$ at day 1. The result is in agreement with a previous study, which reported AOA was inhibited in high N-treated soils [50]. In addition, DMPP + DCD significantly decreased AOA abundance, which indicated that AOA played an important role in ammonia oxidation in the acid soils [51]. In NA, NI treatments significantly inhibited both AOA and AOB at day 1. The result is accordance with a previous study, which reported DMPP significantly
suppressed AOA and AOB in the neutral soil (pH: 7.0) [52]. Moreover, DMPP significantly inhibited AOB abundance during the entire period, which suggests that DMPP was the most effective in neutral soil [48]. In DA, no significant differences in AOA abundance were found between NI treatments and AS during entire incubation time (except day 7), while NIs significantly decreased AOB abundance (Figure 3F, \( p < 0.05 \)), which is in line with previous studies [9,25,41]. These results indicate that AOB plays a critical role in nitrification in alkaline soil as per the study of Jiang et al. [53], which pointed out that AOB dominated in alkaline soil (pH: 8.2).

In addition, higher AOA abundance was found in NI treatments in all tested soils in certain incubation periods (at day 1 in HLJ, day 100 in NA, and day 2 in DA). The results are in contrast to Di and Shi’s studies who showed no impact or reduction in AOA abundance [15,41]. The contrary results indicate that multiple physiochemical properties of soils may affect the efficacy of NIs together [17]. Additionally, our results are in line with a recent study conducted by Fan et al. [54], who reported an increase in AOA abundance with DMPP application. These discrepancies could be due to AOA being promoted by organic compounds [55]. Therefore, NIs might be used as available carbon substrates to stimulate the growth of AOA [17,30]. Furthermore, although AOA and AOB share a common function, their amoA genes are genotypically distinct, indicating the corresponding AMO enzyme may also be physiologically different [56].

5. Conclusions

This study demonstrated that NI treatments could effectively increase soil \( \text{NH}_4^+ \)-N and decrease \( \text{NO}_3^- \)-N concentrations by reducing the soil nitrification process, regardless of soil type and NI type. No significant difference in inhibiting nitrification was detected in HLJ among all NI treatments, while DCD was more efficient in inhibiting nitrification in NA, DMPP + DCD had better efficiency in DA. Therefore, the application of DCD in NA and DMPP + DCD in DA were considered to be the most cost-effective ways to mitigate environmental pollution by reducing \( \text{NO}_3^- \)-N leaching through inhibiting soil nitrification. AOA was dominant in HLJ. NIs treatments significantly inhibited AOB abundance in the two soils with higher pH. Soil pH and soil organic matter were revealed as the main factors in impacting the efficiency of NIs and the abundance of AOA and AOB. Given the inherent limitations of soil incubation experiments, further studies are needed to verify these results under field conditions through covering various soils and climate factors.

**Supplementary Materials:** The following supporting information can be downloaded at: https://www.mdpi.com/article/10.3390/agronomy12020294/s1, Table S1: Inorganic nitrogen concentration of different treatments in three tested soils during the incubation period.

**Author Contributions:** Conceptualization, L.C. and D.L.; formal analysis, L.C.; funding acquisition, D.L. and Z.W.; investigation, D.L.; methodology, L.C.; resources, Y.X., Y.S., F.X., L.Z., P.G. and K.Z.; supervision, L.C., D.L. and Z.W.; writing—original draft, L.C.; writing—review & editing, L.C. and D.L. All authors have read and agreed to the published version of the manuscript.

**Funding:** This research was funded by [the Strategic Priority Research Program of the Chinese Academy of Sciences] grant number [XDA28090200] and [the National Scientific Foundation Project of China] grant number [31971531]. The APC was funded by [the Strategic Priority Research Program of the Chinese Academy of Sciences] grant number [XDA28090200].

**Institutional Review Board Statement:** Not applicable.

**Informed Consent Statement:** Not applicable.

**Data Availability Statement:** All relevant data is contained within the article.

**Conflicts of Interest:** The authors declare no conflict of interest.
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