Variations and Source Apportionment of PM$_{2.5}$ and PM$_{10}$ Before and During COVID-19 Lockdown Phases in Delhi, India

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Abstract: Major cities across the globe including megacity Delhi have experienced considerable lower levels of air pollutants including particulate matter (PM) during COVID-19 lockdown. This study explores pre-lockdown and during lockdown air quality changes in PM$_{2.5}$, PM$_{10}$, PM$_{2.5}$/PM$_{10}$ ratio along with meteorological effects. Selected sites with different pollution signatures in Delhi including Alipur (residential), Okhla (industrial) and Pusa Road (traffic) have experienced mean (S.D.) PM$_{2.5}$ as $87.56(\pm 54.06)$, $124.45(\pm 73.49)$ and $62.14(\pm 58.64) \, \mu g/m^3$ and PM$_{10}$ as $163.01(\pm 77.37)$, $217.71(\pm 93.94)$ and $135.15(\pm 77.90) \, \mu g/m^3$ before lockdown (BL), while for Lockdown 1 (L1), PM$_{2.5}$ concentrations decreased drastically as $39.26(\pm 16.31)$, $38.01(\pm 15.16)$ and $31.03(\pm 12.79) \, \mu g/m^3$ and for PM$_{10}$ as $100.76(\pm 43.71)$, $79.47(\pm 30.97)$ and $66.53(\pm 22.78) \, \mu g/m^3$, respectively, with gradual increase in both pollutants during successive lockdown phase—Lockdown 2, Lockdown 3, Lockdown 4 and Unlock phase 1. The percentage (%) decrease in PM$_{2.5}$ (69.46%) and PM$_{10}$ (63.49%) during lockdown was found well correlated with people mobility (Google and Apple mobility reports), as outdoor activities showed 70–80% decrease in L1 from BL phase. Source apportionment studies suggested both local and regional pollution contribution in Delhi. Comparison of PM$_{2.5}$ and PM$_{10}$ concentrations for the year 2020 with that of 2018 and 2019 and study on diurnal variations of PM$_{2.5}$ and PM$_{10}$ have been discussed.

Keywords: PM$_{2.5}$; PM$_{10}$; PM$_{2.5}$/PM$_{10}$ ratio; COVID-19

1. Introduction

The novel coronavirus outbreak for SARS-CoV-2 variant took place at a very fast rate during COVID-19 pandemic. This variant of coronavirus was firstly detected during December 2019 in Wuhan City of China. COVID-19 was declared as a pandemic event in March 2020 by World Health Organization [1]. On 30 September 2020, global confirmed COVID-19 number of cases was 34,084,559, including 1,016,517 deaths reported from N200 countries/territories worldwide [2]. As per studies reported, six metropolitan cities in India including Delhi, Mumbai, Kolkata, Ahmedabad, Pune and Chennai have known to account, approximately 50% of total reported COVID-19 cases, all over India [3]. COVID-19 lockdown has caused serious reduction in air pollution status across the globe [4–6]. Primary air pollutants (gaseous + particulates) reduction have also been experienced in the Eastern parts of China as a result of adverse effects of meteorological conditions on air pollutants during COVID-19 [7]. As per previous studies, poor air quality of Chinese cities has been found in linkage with higher rate of mortality within those cities [8], whereas lockdown conditions have also been associated with lower pre-mature deaths in some cities as a result of air quality up-gradation during COVID-19 [9, 10]. According to a study in China, a significant relationship has been found between air pollution and infection due to COVID-19 virus among people [11]. In New York, USA, meteorological parameters like average and minimum
temperature, and air quality changes have been known to be linked with effects due to COVID-19 pandemic [12]. According to the reported studies, lower levels of air pollutants were experienced in Barcelona, Spain, during COVID-19 lockdown [13]. In India, metropolitan cities with heavy crowd including Delhi, Mumbai, Kolkata and Bangalore have experienced lower concentrations levels of major air pollutants with significant improvement in their respective air quality during lockdown conditions, in comparison with before lockdown air quality conditions [14, 15].

Ambient atmosphere carries PM of different sizes with their distinct characteristics and causes adverse effects due to their variable chemical compositions, physical characteristics, site location and types of emission sources [16, 17]. Fine particles like PM$_{2.5}$ (particulate matter with diameters less than 2.5 µm) and coarse particles like PM$_{10}$ (particulate matter with diameters less than 10 µm) are most common sizes of particulate matter found in the ambient atmosphere and are of major concern related to research studies. Sources for coarse particles (PM$_{2.5}$ to PM$_{10}$) include re-suspension of loose soil or road dust, natural dust storms and different industrial processes [18], whereas fine particles (PM$_{2.5}$) sources include emissions from heavy traffic activities, energy production processes, biomass burning, etc. [19]. Factors like variable meteorological conditions, land use patterns and population density cause spatio-temporal variations in PM concentrations and other pollutants at both local and regional levels [20, 21].

Since both fine and coarse particles have different physico-chemical properties and diverse sources of emissions/generations, the PM$_{2.5}$/PM$_{10}$ ratio plays a crucial role in providing information related to particulate matter’s origin, their formation processes and thus providing a base to study their associated negative health effects [22, 23]. A higher value for PM$_{2.5}$/PM$_{10}$ ratio shows dominance of fine particles mainly emitted from anthropogenic sources, whereas, smaller PM$_{2.5}$/PM$_{10}$ ratio reveals dominance of coarse particles mainly generated from natural sources including road-dust suspension, natural dust storm, etc. [24]. The values of this ratio show greater spatial variations indicating heterogeneity of PM$_{2.5}$/PM$_{10}$ ratio at different regions as an effect of variable meteorological conditions at different places [25, 26]. Higher PM$_{2.5}$/PM$_{10}$ ratio during winters has been reported [27], due to meteorological parameters including less rainfall or precipitation conditions, low temperature, lower boundary layer depth and stable atmospheric conditions which limit PM$_{2.5}$ dispersion in ambient atmosphere [25, 26, 28].

Higher levels of both PM$_{2.5}$ and PM$_{10}$ found in the ambient atmosphere of Delhi, much more than the standards limits set by the National Ambient Air Quality Standards (NAAQS), India [15, 29]. COVID-19 pandemic in India led to a nationwide lockdown starting from 24 March till 31 May 2020. This nationwide lockdown caused a decreased industrial activities and lesser transportation activities due to which pre-lockdown pollution level of different pollutants, decrease drastically during lockdown event in India [15, 30–33]. Therefore, the present work explores the air quality changes in fine and coarse particulate matter (PM) at different sites of Delhi having different pollution signatures, before lockdown, during different lockdown and unlock phases in Delhi, India.

The objectives of the present study include:

(i) Variations in concentrations of PM$_{2.5}$, PM$_{10}$, PM$_{2.5}$/PM$_{10}$ ratio and meteorological parameters within different site characteristics (Alipur, Okhla and Pusa Road) in Delhi, India, before lockdown and during different phases of lockdown and unlock phases;

(ii) Source apportionment studies to find major reasons for the PM level variations within Delhi, before lockdown and during different phases of lockdown and unlock phases.

(iii) Comparisons of PM (PM$_{2.5}$ and PM$_{10}$) variations (2020) with that of year 2018 and 2019 and comparison of PM diurnal variations (2020) during various lockdown phases (BL, L1, L2, L3, L4, UL1).

2. Materials and Methods

2.1. Site Description

Delhi has been chosen as study area which is the capital city of India and lies at Indo-Gangetic Plains (IGP). Delhi covers an area of ~ 1485 km$^2$ (Coordinates- Latitude: 28°36'36"N; Longitude: 77°13'48"E) (Fig. 1). Delhi is a semi-arid zone having typical IGP climate with four major seasons: pre-monsoon (Mar–May), monsoon (Jun–Sept), post-monsoon (Oct–Nov) and winter (Dec–Feb) seasons. Traffic density, industrial activities, construction works, dust re-suspension, biomass burning, regional transport of pollutants are major contributors of particulate matter (PM$_{2.5}$ and PM$_{10}$) in Delhi, India. Present study includes three representative sampling sites—Alipur as rural, Okhla as Industrial and Pusa Road as Traffic with residential area (Fig. 1).

2.2. Data Collection and Analysis

Real-time data for PM$_{2.5}$ and PM$_{10}$ was collected from air quality monitoring stations (For three sampling sites—Alipur, Okhla and Pusa Road) which have been installed and being monitored in collaboration with CPCB (continuous ambient air quality monitoring, CAAQM and manual
ambient air quality monitoring, MAAQM); DPCC (Delhi Pollution Control Committee); IITM (Indian Institute of Tropical Meteorology), Pune; and, SAFAR (System of Air Quality and Weather Forecasting and Research). The collected data has been converted into daily average data for concentrations ($\mu g/m^3$) of PM$_{2.5}$, PM$_{10}$ and meteorological parameters including wind direction, wind speed, temperature and relative humidity. The raw data used here, is available at CPCB online portal for air quality data dissemination [34] which has been analysed for air quality assessment studies for before and during the different lockdown periods of COVID-19 and unlock phases. The data follow quality assurance or quality control (QA/QC) protocols which is done by CPCB using rigorous protocols
Table 1 Mean values of relative humidity, air temperature, wind speed, PM$_{2.5}$, PM$_{10}$ and PM$_{2.5}$/PM$_{10}$ during different lockdown and unlock phases (BL, L1, L2, L3, L4 and UL1) in Delhi

| Sampling locations | Phases/ events | Mean relative humidity (RH %) | Mean atmospheric temperature (AT in °C) | Mean wind speed (WS in m/s) | Mean PM$_{2.5}$ (ug/m$^3$) | Mean PM$_{10}$ (ug/m$^3$) | Mean PM$_{2.5}$/PM$_{10}$ ratio |
|--------------------|---------------|-------------------------------|----------------------------------------|----------------------------|---------------------------|---------------------------|-------------------------------|
| Alipur             | Before lockdown (BL) | 61.17 (± 5.92) | 18.48 (± 3.97) | 1.16 (± 0.05) | 87.56 (± 54.06) | 163.01 (± 77.37) | 0.54 (± 0.10) |
| Okhla              | Before lockdown (BL) | 66.57 (± 11.87) | 17.54 (± 3.72) | 0.82 (± 0.32) | 124.45 (± 73.49) | 217.71 (± 93.94) | 0.54 (± 0.11) |
| Pusa Road          | Before lockdown (BL) | 58.10 (± 9.16) | 23.52 (± 3.96) | 1.32 (± 1.00) | 62.14 (± 58.64) | 135.15 (± 77.90) | 0.48 (± 0.10) |
| Alipur             | Lockdown 1 (L1) | 50.08 (± 7.34) | 24.94 (± 2.78) | 1.16 (± 0.05) | 39.26 (± 16.31) | 100.76 (± 43.71) | 0.40 (± 0.08) |
| Okhla              | Lockdown 1 (L1) | 42.05 (± 11.89) | 27.21 (± 2.77) | 0.86 (± 0.27) | 38.01 (± 15.16) | 79.47 (± 30.97) | 0.48 (± 0.07) |
| Pusa Road          | Lockdown 2 (L2) | 51.45 (± 8.92) | 25.85 (± 2.57) | 1.89 (± 0.68) | 31.03 (± 12.79) | 66.53 (± 22.78) | 0.47 (± 0.12) |
| Alipur             | Lockdown 2 (L2) | 44.99 (± 5.69) | 28.66 (± 1.78) | 1.16 (± 0.04) | 50.72 (± 18.36) | 147.24 (± 47.69) | 0.34 (± 0.08) |
| Okhla              | Lockdown 2 (L2) | 41.36 (± 11.93) | 29.87 (± 2.09) | 0.75 (± 0.17) | 43.15 (± 14.54) | 107.92 (± 36.22) | 0.40 (± 0.09) |
| Pusa Road          | Lockdown 3 (L3) | 46.77 (± 7.81) | 28.39 (± 3.46) | 1.20 (± 0.67) | 49.35 (± 36.57) | 101.92 (± 54.53) | 0.44 (± 0.12) |
| Alipur             | Lockdown 3 (L3) | 45.63 (± 4.64) | 30.12 (± 1.55) | 1.23 (± 0.03) | 65.42 (± 20.45) | 136.12 (± 42.97) | 0.48 (± 0.04) |
| Okhla              | Lockdown 3 (L3) | 41.70 (± 10.13) | 31.33 (± 2.04) | 0.77 (± 0.18) | 46.42 (± 14.30) | 114.37 (± 30.96) | 0.40 (± 0.05) |
| Pusa Road          | Lockdown 4 (L4) | 46.86 (± 8.39) | 31.20 (± 3.12) | 2.21 (± 1.42) | 57.01 (± 14.02) | 115.64 (± 21.57) | 0.49 (± 0.09) |
| Alipur             | Lockdown 4 (L4) | 38.40 (± 10.69) | 33.05 (± 3.51) | 1.23 (± 0.03) | 68.43 (± 35.12) | 180.90 (± 76.68) | 0.39 (± 0.13) |
| Okhla              | Lockdown 4 (L4) | 30.69 (± 21.50) | 34.87 (± 4.21) | 0.90 (± 0.30) | 47.46 (± 24.40) | 153.21 (± 62.84) | 0.32 (± 0.11) |
| Pusa Road          | Lockdown 4 (L4) | 41.24 (± 12.95) | 34.00 (± 3.82) | 2.58 (± 1.26) | 46.03 (± 23.97) | 132.10 (± 51.52) | 0.36 (± 0.12) |
| Alipur             | Unlock 1 (UL1) | 51.18 (± 3.79) | 32.32 (± 2.82) | 1.27 (± 0.09) | 51.05 (± 11.58) | 120.16 (± 36.60) | 0.36 (± 0.08) |
| Okhla              | Unlock 1 (UL1) | 54.72 (± 5.72) | 32.33 (± 2.54) | 0.54 (± 0.13) | 46.66 (± 10.63) | 135.59 (± 45.92) | 0.42 (± 0.08) |
| Pusa Road          | Unlock 1 (UL1) | 51.27 (± 4.51) | 33.43 (± 2.77) | 1.83 (± 0.51) | 39.10 (± 10.52) | 94.70 (± 30.55) | 0.44 (± 0.06) |

for the calibration of the instruments. Data from January to June 2020 have been studied for analysing air quality before and during different COVID-19 lockdown and unlock phases as mentioned below:

Different phases of lockdown and unlock during COVID-19

Before lockdown (BL): 1 January–24 March 2020 (~3 months);
Lockdown phase-1 (L1): 25 March–14 April 2020 (21 days);
Lockdown phase-2 (L2): 15 April–3 May 2020 (19 days)

Lockdown phase-3 (L3): 4 May–17 May 2020 (14 days);
Lockdown phase-4 (L4): 18 May–31 May 2020 (14 days);
Unlock phase-1.0 (UL1): 1 June–30 June 2020 (30 days; data analysis done for 20 days).

Percentage (%) change is calculated by using following formula (e.g. % change for BL–L1):

\[(BL – L1) / BL \times 100\].

Here, percentage (%) change is calculated for L1 with respect to BL phase.
Data for human mobility trends have been obtained from Google Mobility Reports and Apple Mobility Reports. Wind Trajectories have been obtained using NOAA HYSPLIT Trajectory model [35], and fire count data have been obtained using NASA’s MODIS active fire data ([36]; Figure Courtesy-[37]). NASA’s Fire Information for Resource Management System (FIRMS) used here provides fire maps for specific dates on global location to confirm biomass burning activities. NOAA HYSPLIT Trajectory Model used here provides maps for back trajectories for wind which shows contribution of local and regional polluting sources at a specific location.

3. Results and Discussion

3.1. Measurements of Surface Meteorology

Variations in surface meteorology for meteorological parameters including relative humidity (%), atmospheric temperature (°C) and wind speed (m/s) are provided in Table 1 for all the three sampling sites for different phases of BL, L1, L2, L3, L4 and UL1. Highest mean relative humidity (R.H.) is observed during before lockdown phase whereas lowest during L4 phase at all the three sampling sites. Air temperature (°C) is found lowest during before lockdown phase and highest during L4 phase. The reason behind this is average values of before lockdown phase also consists of winter season values whereas other lockdown phases and L4 have predominantly summer season with higher temperature and lower R.H. values. At both Okhla and Pusa Road, highest average wind speed has been observed during L4 phase whereas for Alipur it is observed during UL1 phase. Higher wind speed helps in dispersion of pollutants and lower wind speed causes accumulation of pollutants within a specific area.

It is reported that lower temperature plays a major role in pollution build-up due to lower boundary layer during winters while higher relative humidity during winters leads to secondary aerosol particle formation, increasing the PM concentrations [38, 39]. Recent studies have revealed that cold and dry conditions were found to accelerate the rate of coronavirus spread [40–46] and decrease in relative humidity leads to an increase in viral spread [47–49]. The coronavirus is found less airborne active in hot and humid conditions [50]. Therefore, the climatic condition with lower R.H. in Delhi, has more transmission potential for the coronavirus spread among residents and during winters, the COVID-19 virus spread is presumed to spread with faster rate due to colder temperature, whereas the virus spread will be lower during hot and dry conditions of summer.

3.2. Variations in PM2.5 and PM10 Before and During COVID-19 Lockdown

Table 1 shows mean values for PM2.5 and PM10 concentrations. Highest mean values for concentrations (µg/m³) were reported for both PM2.5 and PM10, before COVID-19 lockdown phase while lowest values were reported during L1 Phase, at all the three sampling sites. The reason for the difference in these concentrations’ values is due to higher emissions of PM before COVID-19 lockdown and sudden decrease in emitting sources due to COVID-lockdown conditions. For pre-COVID concentrations, Okhla reported highest values for PM2.5 and PM10 due to industrial emissions, followed by Alipur due to vehicular emissions at national highway and biomass burning activities in nearby agricultural fields and landfill waste burning, and then Pusa Road due to traffic emissions. Similar studies have shown effect of COVID-19 lockdown on the reduced concentrations of air pollutants in Delhi and other parts of India [30, 31, 33].

After L1 phase, both PM2.5 and PM10 concentrations found gradual increasing for L2, L3 and L4 phases as slight relaxation of lockdown measures was observed after L1 phase (from 7 April 2020) for both vehicular movement and industrial processes and functioning, beyond the red zone (government recognized infected zone from major COVID-19 cases) (Figs. 2, 3). For unlock phase (UL1), due to rainy events occurred in monsoon season, concentrations for both pollutants decreased even after visible movements of residents. PM2.5 and PM10 have reduced below the
NAAQS permissible limit set by CPCB, at all the three sites during L1 and increased for some days during L2, L3, L4 and UL1 (Figs. 2, 3). However, PM$_{2.5}$ and PM$_{10}$ concentrations were found reduced below WHO standard limit on very few days during lockdown phases (Figs. 2, 3). The present study showed mean concentrations (Standard deviations; S.D.) at all the three selected sampling sites—Alipur, Okhla, Pusa, for PM$_{2.5}$ as 87.56 (± 54.06), 124.45 (± 73.49) and 62.14 (± 58.64) l/g/m$^3$ and for PM$_{10}$ as 163.01 (± 77.33), 217.71 (± 93.94) and 135.15 (± 77.90) l/g/m$^3$ before lockdown (BL), while both PM$_{2.5}$ and PM$_{10}$ concentrations were drastically decreased during (Lockdown-1 phase) L1 (Table 1). For L1, PM$_{2.5}$ concentrations were reported as 39.26 (± 16.31), 38.01 (± 15.16) and
31.03 (± 12.79) µg/m³ while for PM10 as 100.76 (± 43.71), 79.47 (± 30.97) and 66.53 (± 22.78) µg/m³ with gradual increase in concentrations for both pollutants at all the three sampling sites during successive lockdown phases including L2, L3, L4 and unlock phase—UL1 (Table 1; Figs. 2, 3). PM2.5 percentage (%) change from BL–L1, L1–L2, L2–L3, L3–L4, L4–UL1 is calculated as −55.16, 29.19, 28.15, 5.28, −25.39 for Alipur, −69.46, 13.52, 7.58, 2.24, −1.68 for Okhla and −59.04, 15.52, −19.26, −15.27 for Pusa Road, respectively. Also, PM10 percentage (%) change from BL–L1, L1–L2, L2–L3, L3–L4, L4–UL1 is calculated as −38.19, 46.13, −7.55, 32.90, −33.58 for Alipur, −63.50, 35.80, 5.98, 33.96, 14.50 for Okhla and −50.77, 53.19, 14.46, 12.15, −8.26 for Pusa Road site, respectively. Srivastava et al. [31] have reported percentage (%) decrease of ~58% and 47% in PM10 and PM2.5 concentration, respectively, in Delhi. Highest percentage (%) decrease at Okhla industrial site than other sites during lockdown period shows lesser industrial activities during the period whereas almost 50% decrease in PM concentrations at Pusa Road may be due to lesser traffic activities during lockdown period and percentage (%) decrease in PM concentrations, particularly PM10, at Alipur may be attributed to somewhat lower but ongoing rural activities, landfill burnings, agricultural burnings but decrease in vehicular emissions at local roads and nearby national highways. Percentage (%) change of PM for L1, L2, L3, L4 and UL1 conditions with that of BL is shown in Table 2. This table also confirms the above discussions along with the fact that at Alipur site percentage (%) decrease of 55.16% in BL–L1 shows half cut in PM2.5 sources as an effect of sudden lockdown conditions while −10.97% increase in PM10 in BL–L4 shows effect of dust-storms and biomass burning activities. More than half cut (> 60% decrease) in PM2.5 emissions is attributed to lower/closed industrial activities as Okhla site. At Pusa Road site percentage (%) decrease in PM2.5 and PM10 in BL–L1 shows the effect of lockdown activities with lower traffic emission and road-dust suspension and lesser % decrease in PM2.5 by only 8.26% in BL–L3 shows effect of biomass burning activities and lenient lockdown rules. PM10 concentrations for Alipur also remained highest among other sites, during the whole lockdown period despite being rural area which may be due to the mentioned ongoing localized activities like agricultural burning, landfill burning, road-dust suspension, dust-storm events, etc. (Fig. 3). Higher PM2.5 concentrations values at rural area of Alipur during L3 and L4 and fire count data confirms crop burning events at localized and regional areas nearby Delhi (Figs. 2, 10).

According to human mobility trend reports including Google mobility reports [51] and Apple mobility reports [52] (Figs. 4, 5), % increase in people mobility trend showed ~40% increase during L1, L2, L3 and ~30% increase during L4 and UL1, in residential activities with respect to before lockdown phase. Outdoor activities like visiting grocery stores, retails, parks, transit stations and workplaces have decreased during L1 (70–80%) and gradually increased in successive phases of lockdown and unlock phases with up to ~20% t during UL1 phase (Figs. 4, 5). This shows majority of Delhi residents were inside their homes during lockdown phases while some residents went to buy essentials and groceries, personal works and in transit for meeting their family members living nearby or migrating to places. Also, with the successive lockdown phases, outdoor activities increased but significant number of people were found to be associated with residential activities. The gradual increase in both PM2.5 and PM10 concentrations was found well correlated with people mobility during different lockdown phases.
Similar trends for decrease in air pollutants particularly, PM$_{2.5}$ and PM$_{10}$, were found associated with human mobility trend in Singapore, during COVID-19 lockdown [53]. Since outdoor pollution concentrations also affect indoor pollution, people living inside their houses were also exposed to outdoor concentrations along with those going outdoors. Studies confirmed that outdoor PM concentrations can easily enter buildings and vehicles affecting their respective indoor quality [54, 55]. According to a research carried out in Germany, both outdoor and indoor PM$_{2.5}$ levels were found well correlated with each other with a significant correlation coefficient ($r = 0.82$) whereas, similar results for correlations of outdoor and indoor PM were observed in Guangzhou and Beijing City of China [56–58]. Studies reported that indoor PM concentrations were found similar to the outdoor ones when the influence of other indoor sources is minimum. Modelling studies of PM$_{2.5}$ suggested that absence of indoor sources, may be linked with the 50–70% presence of the outdoor PM$_{2.5}$ concentrations at indoors [55, 59]. According to an Oxford study, in the presence of indoor activities like cleaning, cooking or smoking, PM concentrations exposure can even exceed than that of outdoor concentrations exposure [60]. Indoor sources like cleaning, dusting, walking, use of other domestic or office equipment, painting, smoking, etc. can increase PM concentrations at indoors [61, 62]. According to study of Birmingham, Wales and Cornwall, high concentrations of PM$_{10}$ found indoors with their chemical composition also affected by the sources which were present outdoors [63, 64]. Therefore, living indoors is also affected by outdoor concentrations of pollutants depending on meteorological conditions, ventilation systems present in the house and specific polluting sources at indoors.

Also, poor ventilation system and super-spreader events like large gatherings at indoors, may act as an enhancer for increase in COVID cases with increase in fine PM acting as a carrier for droplet transport. If indoor is heavily crowded with people along with poor air circulation in building, increase in their physical activity like loud talking, playing, laughing, singing and dancing can cause higher breathing rate, thus, increasing the number of micro-droplets release leading to high spread rate of viruses via respiratory droplets [3, 65]. According to Kay [66], super-spreader events play a major role in the faster spread of coronavirus and he has made an international database for Indian cities in a comprehensive way. The database consists of a list of major super-spreading events including large clusters for COVID-19 infection for a period including February and March 2020. The study suggested that most of the coronavirus infections outbreak were linked to indoor conditions where people were closely packed at the places like home, social gathering, workplace, public transport and restaurants [66]. According to another study, top 50 coronavirus outbreaks occurred at the large gatherings including events and places like weddings, funerals, religious places, prison, call centres, food packing centres, networking events related to business, etc. [3].

![Mobility Trends](image)

**Fig. 5** Mobility report for Delhi, India, from 15.02.2020 to 20.06.2020 for different places such as retail, grocery stores, parks, transit stations, workplaces and residential complexes. (Apple Mobility Reports, [52])
3.3. Variation in PM$_{2.5}$/PM$_{10}$ Ratio Before and During COVID-19 Lockdown

PM$_{2.5}$ and PM$_{10}$ have different sources of emissions/generation, therefore, PM$_{2.5}$/PM$_{10}$ ratio shows different characteristics of particle pollution, although, the ratio greatly varies at both spatial and temporal scale due to time and site-specific changes in PM concentrations. Mean values of PM$_{2.5}$/PM$_{10}$ ratio are given in Table 1. The present study shows PM$_{2.5}$/PM$_{10}$ ratio before lockdown as 0.54 ($\pm$ 0.10, $\pm$ 0.11) for both Alipur and Okhla sites and 0.48 ($\pm$ 0.10) for Pusa Road site, while L1 showed PM$_{2.5}$/PM$_{10}$ ratio as 0.40 ($\pm$ 0.08) for Alipur, 0.48 ($\pm$ 0.07) for Okhla and 0.47 ($\pm$ 0.12) for Pusa Road, respectively (Table 1). Before lockdown, for Pusa Road and Okhla PM$_{2.5}$/PM$_{10}$ ratio found > 0.5 showing dominance of fine-sized particles with higher contribution from anthropogenic emissions, while at Alipur, the ratio was 0.48 ($\pm$ 0.10) showing dominance of coarse-sized particles than the fine particles (Table 1). At all the three sites fine mode particles were mostly dominant before lockdown condition with PM$_{2.5}$/PM$_{10}$ ratio up to 0.80 while during lockdown coarse mode particles become dominant (Fig. 6). Higher ratios of PM$_{2.5}$/PM$_{10}$ show contribution of anthropogenic sources in the particle pollution, whereas, smaller PM$_{2.5}$/PM$_{10}$ ratios show presence of more coarse-sized particles in particle pollution, majorly emitted/generated from natural sources like dust storm [24]. Higher PM$_{2.5}$/PM$_{10}$ ratio in Delhi, before lockdown, at Okhla site is attributed to nearby Industrial activities, at Pusa Road it is due to higher traffic density and at Alipur site, it is due to localized agricultural and landfill burning activities along with more or less vehicular emissions at all the three sites. Highest variation in PM$_{2.5}$/PM$_{10}$ ratio observed at Okhla and Alipur sites due to presence of more complex and changing PM sources at these sites than at the Pusa Road site. According to a study conducted in three east-central US states, fine particles (PM$_{2.5}$) are found to contribute 67% of coarse particles (PM$_{10}$) whereas, study on 20 European areas showed average PM$_{2.5}$/PM$_{10}$ ratio as 0.60 [67, 68]. In Saudi Arabia, average PM$_{2.5}$/PM$_{10}$ ratio is found as 0.33 mainly due to contribution of coarse particles (sand/dust) from desert area [69]. According to a study conducted in the USA, higher PM$_{2.5}$/PM$_{10}$ ratios were observed in the eastern (~ 0.7) part than the central or western (~ 0.5) parts of the USA [70]. Urban sites of Wuhan, China showed highest ratio as 0.75 during winter and lowest in summer as 0.55 [71]. Higher PM$_{2.5}$/PM$_{10}$ ratio reported during winter or autumn than in summer or spring [71, 72] mainly due to increase in fine particles emissions or secondary aerosol formation due to higher fuel consumption for domestic and industrial heating and lower mixing height, during winters [38, 73]. Also, stable atmospheric conditions during winter cause wet and dry deposition of aerosols which favours accumulation of fine particles in the atmosphere due to which fine particles become dominant in PM$_{10}$ during winter.
[39]. PM$_{2.5}$/PM$_{10}$ percentage (%) change from BL–L1, L1–L2, L2–L3, L3–L4, L4–UL1 is calculated as $-25.93$, $-15$, $-18.75$, $-7.69$ for Alipur, $-11.11$, $-16.67$, $0$, $-20$, $31.25$ for Okhla and $-2.08$, $-6.38$, $11.36$, $-26.53$, $22.22$ for Pusa Road site, respectively.

3.4. Source Apportionment of PM (PM$_{2.5}$ and PM$_{10}$) at Selected Sites—Alipur, Okhla and Pusa Road Before and During COVID-19 Lockdown

Polar Plots of PM$_{2.5}$ concentrations have been plotted against wind speed and wind direction for BL, L1, L2, L3, L4, UL1.
L4 and UL1 phases at all the three sampling sites (Figs. 7, 8, 9). Alipur, during BL phase, most of the higher concentrations were reported to coming from W and NW directions with wind speed varying between (0–0.5 m/s) showing dominance of local contribution of PM$_{2.5}$ sources mainly from vehicular emissions or local biomass burning activities (Fig. 7). Some higher PM$_{2.5}$ emissions were also coming from N, S and SW directions with higher wind speed up to 1.25 m/s which shows both local and regional transfer of PM$_{2.5}$ at Alipur site. During L1, most of the higher concentration were associated with S, SW, W and NW with wind speed 0–1.25 m/s. Higher PM$_{2.5}$ concentrations were found with, SW and S directions with 0.25–1.25 m/s for L2; SW with 0.75 to 1 m/s for L3; S...

Fig. 8 Polar plots of the hourly variations in wind speed (round radius, in units of m/s) and direction (angles) to surface PM$_{2.5}$ concentrations (colour contours, in units of μg/m$^3$) at Okhla, Delhi a from 01.01.2020 to 24.03.2020 (Before Lockdown, BL), b from 25.03.2020 to 14.04.2020 (Lockdown 1, L1), c from 15.04.2020 to 03.05.2020 (Lockdown 2, L2), d from 04.05.2020 to 17.05.2020 (Lockdown 3, L3), e from 18.05.2020 to 31.05.2020 (Lockdown 4, L4), f from 01.06.2020 to 20.06.2020 (Unlock 1, UL1)
with 1–1.25 m/s for L3; and, SE with 0.25–0.75 m/s. Therefore, at Alipur, most of the higher concentrations were associated with S, SW, N, NW with wind speed varying between 0.25 and 1.25 m/s showing contribution of both local and regional PM$_{2.5}$ sources (Fig. 7).

At Okhla, BL phase has higher PM$_{2.5}$ Contribution from N and NE direction with lower wind speed 0–1 m/s showing effect of local emission sources like nearby Okhla industrial estate (Fig. 8). Higher PM$_{2.5}$ concentrations were found associated with, W with 0–0.25 m/s wind speed during L1; W and NW with 0–0.25 m/s wind speed during L2; W, NW and SW with 0–1 m/s wind speed during L3; W and SW with < 0.5 m/s wind speed during L4; and, S with 0.25–0.5 m/s wind speed during UL1 phase. PM$_{2.5}$ pollution at Okhla was found mostly associated with lower wind speed 0–1 m/s confirms the local source contribution towards PM$_{2.5}$ during different phases of study (Fig. 8).

At Pusa Road, higher PM$_{2.5}$ concentrations were reported for E direction with 0–1 m/s wind speed during BL phase; W with 0–1 m/s wind speed during L1; W and NW with 0–0.25 m/s wind speed during L2; W, NW and SW with 0–1 m/s wind speed during L3; W and SW with < 0.5 m/s wind speed during L4; and, S with 0.25–0.5 m/s wind speed during UL1 phase. PM$_{2.5}$ concentrations were reported for E direction with 0–1 m/s wind speed during BL phase; W with 0–1 m/s wind speed during L1; S with

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**Fig. 9** Polar plots of the hourly variations in wind speed (round radius, in units of m/s) and direction (angles) to surface PM$_{2.5}$ concentrations (colour contours, in units of gg/m$^3$) at Pusa Road, Delhi a from 01.01.2020 to 24.03.2020 (Before Lockdown, BL), b from 25.03.2020 to 14.04.2020 (Lockdown 1, L1), c from 15.04.2020 to 03.05.2020 (Lockdown 2, L2), d from 04.05.2020 to 17.05.2020 (Lockdown 3, L3), e from 18.05.2020 to 31.05.2020 (Lockdown 4, L4), f from 01.06.2020 to 20.06.2020 (Unlock 1, UL1)
0–0.5 m/s wind speed during L2; E and NE with 0–1 m/s wind speed during L3; E and NE with 0–1 m/s wind speed phase; and, S and SE with 2–3 m/s wind speed during UL1 phase (Fig. 9). Highest wind speeds up to 5.5 m/s has been reported at Pusa Road site mostly during L3 and L4 phase showing effect of regional contribution at this site during the mentioned lockdown period (Fig. 9). The strong near-surface wind causes the intensification of PM 2.5 during heavy air pollution periods [74].

Fire count data plots show major crop residue burning (CRB) events across Northern India during May 2020, i.e. during L3 and L4 phases (Fig. 10). Higher PM2.5 concentration at Alipur rural site during L3 and L4 confirms the effect of CRB events in Delhi (Fig. 2). Figure 11 shows NOAA HYSPLIT wind trajectory plots for all the three sites, from 09.05.2020 to 15.05.2020 (for L3 phase) duration. These wind trajectory plots confirmed the pollutants from fire events and wind during L3 phase found to be coming from areas where major crop burning (CRB) activities took place across Northern India thus affecting selected sampling sites for PM2.5 pollution by regional transfer of pollutants from biomass burning activities during lockdown phases L3 and L4. The meteorological conditions of ambient atmosphere greatly affect atmospheric processes like transport, diffusion, dispersion, transformations and depositions of PM present in the atmosphere. The results of source apportionment studies confirm that wind direction in Delhi is a major factor with affects PM concentrations and acts as an indicator of the natural and anthropogenic sources and their locations present in the specific directions. Wind direction also affects temporal and seasonal variations of PM concentrations [75, 76].
3.5. Comparison of PM Variations with that of 2018 and 2019

Figure 12 shows daily average concentrations (µg/m³) of PM$_{2.5}$ and PM$_{10}$ for three different sites—Alipur (fig: a and b), Okhla (fig: c and d), Pusa Road (fig: e and f), respectively, before, during and after lockdown phases (year 2020) and comparisons with that of the years 2018, 2019. The study suggests that 24 h average PM$_{2.5}$ and PM$_{10}$ concentrations are minimum for the year 2020 during various lockdown phases whiles the same is higher for 2018 and 2019 for the same period without lockdown conditions. This shows, in general, all the sampling sites chosen are greatly affected by the specific sources of air pollution without lockdown conditions as discussed in previous sections of this paper. Similar comparison studies of lockdown period with pre-lockdown conditions for years 2019 and 2020 have been done by Chauhan and Singh [30] and Zhang et al. [33].

3.6. Comparison of PM Diurnal Variations During Various Lockdown Phases

Figure 13 shows diurnal variations of hourly average concentrations (µg/m³) of PM$_{2.5}$ and PM$_{10}$ for three different sites—Alipur (fig: a and b), Okhla (fig: c and d), Pusa Road (fig: e and f), respectively, before, during and after lockdown phases including BL, L1, L2, L3, L4, UL1 phases (year 2020). The comparison study confirms that Alipur site at 6.00 h of L4 shows highest average
concentration for both PM$_{2.5}$ and PM$_{10}$ may be due to biomass burning activities. Okhla site is characterized by highest PM$_{2.5}$ and PM$_{10}$ at 8.00 h for BL phase attributed by industrial activities during before lockdown (BL) conditions. Pusa Road site shows highest concentrations of both PM$_{2.5}$ and PM$_{10}$ during 2.00 h in L3 phase mainly due to haze events. Haze events were found to dominate Delhi region in L3 phase during night-time which increased PM$_{2.5}$ concentration as reported by Dhaka et al. [77].

### 4. Summary and Conclusion

The findings of the study may be summarized as:

- The present study showed higher concentrations of both PM$_{2.5}$ and PM$_{10}$ concentrations at all the three selected sampling sites of Delhi—Alipur, Okhla, Pusa before lockdown (BL), which drastically decreased during Lockdown-1 (L1) phase. Gradual increase in concentrations for both pollutants at all the three sampling
sites observed during successive lockdown phases including L2, L3, L4 and unlock phase—UL1. Highest % decrease at Okhla industrial site than other sites shows lesser industrial activities at this site during lockdown period whereas almost 50% decrease in PM concentrations at Pusa Road may be due to lesser traffic activities during lockdown period and lower % decrease in PM concentrations at Alipur may be attributed to lower but ongoing localized activities like agricultural burning, landfill burnings and lower emissions from vehicular emissions. At all the three sites fine mode particles were mostly dominant before lockdown condition with PM$_{2.5}$/PM$_{10}$ ratio up to 0.80 while during lockdown coarse mode particles become dominant. Comparison of PM$_{2.5}$ and PM$_{10}$ concentrations for the year 2020 with that of 2018 and 2019 and study on diurnal variations of PM$_{2.5}$ and PM$_{10}$ also confirmed the discussed emission sources.

- The gradual decrease/increase in concentrations of both PM$_{2.5}$ and PM$_{10}$ were found well correlated with people mobility during successive lockdown phases. According to Google and Apple mobility reports, outdoor activities like visiting grocery stores, retails, parks, transit stations and workplaces have decreased during L1 (70–80%) and gradually increased in successive phases of lockdown and unlock phases with up to −20% during UL1 phase.
- Polar Plots of PM$_{2.5}$ concentrations against wind speed and wind showed that specific wind directions and wind speeds are associated with most of the higher concentrations at all the three sampling sites. Fire count data plots showed major crop residue burning (CRB) events across Northern India during May 2020, i.e. during L3 and L4 phases. NOAA HYSPLIT wind trajectory plots for all the three sites also confirmed that the wind during L3 phase found to be coming from areas where
CRB activities took place across Northern India thus affecting selected sampling sites for PM$_{2.5}$ pollution by regional transfer of pollutants from biomass burning activities during lockdown phases L3 and L4.

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