Pesticide Urinary Metabolite Levels of Children in Eastern North Carolina Farmworker Households

Thomas A. Arcury, Joseph G. Grzywacz, Dana B. Barr, Janeth Tapia, Haiying Chen and Sara A. Quandt

doi:10.1289/ehp.9975 (available at http://dx.doi.org/)
Online 28 March 2007
Pesticide Urinary Metabolite Levels of Children in Eastern North Carolina Farmworker Households

Thomas A. Arcury¹
Joseph G. Grzywacz¹
Dana B. Barr²
Janeth Tapia³
Haiying Chen⁴
Sara A. Quandt⁵

¹Department of Family and Community Medicine, Wake Forest University School of Medicine, Winston-Salem, North Carolina

²National Center for Environmental Health, Centers for Disease Control and Prevention, Atlanta, Georgia

³North Carolina Farmworkers Project, Benson, North Carolina

⁴Department of Biostatistical Sciences, Division of Public Health Sciences, Wake Forest University School of Medicine, Winston-Salem, North Carolina

⁵Department of Epidemiology and Prevention, Division of Public Health Sciences, Wake Forest University School of Medicine, Winston-Salem, North Carolina

Correspondence Thomas A. Arcury, PhD
Department of Family and Community Medicine
Wake Forest University School of Medicine
Medical Center Boulevard
Winston-Salem, NC 27157-1084
Phone: 336-716-9438
Fax: 336-716-3206
E-mail: tarcury@wfubmc.edu
Running Title: Farmworker Child Pesticide Metabolite Levels

Article Descriptor: Children’s Health

Keywords: agriculture, biomonitoring, child health, farmworker, health disparities, Latino/Hispanic, occupational health, pesticide exposure, pesticide metabolites

Acknowledgements: The research reported in this paper was supported by grant R25 OH07611 from the National Institute for Occupational Safety and Health. Arcury has received funding for the development of pesticide safety educational materials from Syngenta Crop Protection, Inc. The other authors declare they have no competing financial interests.

Abbreviations

2,4,5-T  2,4,5-trichlorophenoxyacetic acid
2,4-D  2,4-Dichlorophenoxyacetic acid
3PBA  3-phenoxybenzoic acid
ACE  acetochlor mercapturate
ATZ  atrazine mercapturate
C  centigrade
CDC  Centers for Disease Control and Prevention
CIT  5-chloro-1,2-dihydro-1-isopropyl-[3H]-1,2,4-triazol-3-one
CMHC  3-chloro-4-methyl-7-hydroxycoumarin
DEAMPY  2-diethylamino-6-methyl pyrimidin-4-ol
DEET  diethyl-m-toluamide
GM  geometric mean
IMPY  2-isopropyl-4-methyl-6-hydroxypyrimidinol
LOD  Limit of detection
MDA  malathion dicarboxylic acid
MET  metolachlor mercapturate
NHANES  National Health and Nutrition Examination Survey
OP  organophosphorous
PNP  para-nitrophenol
TCPY  3,5,6-trichloro-2-pyridinol
U.S. EPA  United States Environmental Protection Agency
U.S. GAO  United States Government Accountability Office
Outline

Abstract

Introduction

Materials and Methods
   Sample and Data Collection
   Laboratory Analysis
   Measures
   Data Analysis

Results
   Exposure, Safety Behavior, and Personal Characteristics
   Pesticide Metabolite Levels
   Pesticides Detected per Child:  Number and Predictors

Discussion

References

Table 1: Selected Agricultural Characteristics of Counties from which Participants Were Recruited, 2002 Census of Agriculture.

Table 2: Exposure Characteristics of Children Aged 1 to 6 Years Living in Farmworker Families, Eastern North Carolina, 2004.

Table 3: Pesticide Urinary Metabolites for Children Age 1 to 6 Years (n=60) in Eastern North Carolina Farmworker Households, 2004:  Number of Detects, Median and Geometric Mean, and Creatinine Adjusted Median and Geometric Mean.

Table 4: Predictors with a Significant Relationship to Number of Pesticide Urinary Metabolite Detects, Children Age 1 to 6 Years (n=60) in Eastern North Carolina Farmworker Households, 2004.

Table 5: Comparison of Percent Detects, Creatinine Unadjusted and Adjusted Geometric Mean, 50th Percentile, and 95th Percentile, for Children Aged 1 – 6 Years (n=60) in Eastern North Carolina Farmworker Households (Age 1-6), 2004, and 1999-2000 NHANESa Data for Total, Mexican-American, and Children Aged 6 – 11 Years.
Abstract

**Background:** This investigation documents the pesticide urinary metabolite levels of farmworker children in North Carolina; determines the number of different metabolites detected for each child; and delineates risk factors associated with the number of metabolites.

**Methods:** Urine samples were collected from 60 Latino farmworker children aged 1-6 years (34 female, 26 male). Interviews were completed by their mothers in Spanish. Urine samples were analyzed for 14 pesticide metabolites, including the organophosphate pesticides chlorpyrifos, coumaphos, diazinon, isazaphos, malathion, pirimiphos, and parathion and its methyl counterpart; a common metabolite of at least 18 pyrethroid insecticides; the repellent DEET; and the herbicides 2,4,5-trichlorophenoxyacetic acid, 2,4-dichlorophenoxyacetic acid, acetochlor, atrazine and metolachlor. Predictors included measures of para-occupational, residential and environmental exposure, child characteristics, and mother characteristics.

**Results:** 13 metabolites were present in the urine samples. Organophosphate pesticide metabolites were detected in a substantial proportion of children, particularly metabolites of parathion/methyl parathion (PNP) (90.0%; geometric mean 1.00 µg/L), chlorpyrifos/chlorpyrifos methyl (TCPY) (83.3%; geometric mean 1.92 µg/L), and diazinon (IMPY) (55.0%; geometric mean 10.56 µg/L). The number of metabolites detected ranged from 0-7, with a mode of 4 detected (28.3%). Boys, those living in rented housing, and those with mothers working part-time had more metabolites detected.

**Conclusions:** Children in farmworker homes experience multiple sources of pesticide exposure. Pesticides may remain in their environments for long periods. Environmental and occupational health changes are needed to address these exposures. Research is needed with more precise measures of exposure and on the health effects of concurrent exposure to multiple pesticides.
Introduction

Farmworker children, as well as their parents, are exposed to pesticides (Arcury et al. 2005, 2006; Coronado et al. 2004; Lambert et al. 2005; Quandt 2004; Strong et al. 2004; Thompson et al. 2003). Pesticide exposure has health consequences for all persons exposed (Reigart and Roberts 1999). Immediate consequences of limited pesticide exposure include rash, nausea, vomiting, and blurry vision. Immediate effects of significant exposure include disorientation, loss of continence, coma, and death. Delayed consequences of limited or significant exposure may include sterility, birth defects, neurodegenerative disease, and cancer (Reigart and Roberts 1999).

The potential consequences of pesticide exposure are greater for children than for adults (Eskenazi et al. 1999; Faustman et al. 2000; Weiss et al. 2004). Due to short stature and characteristic behaviors, children have greater exposure to pesticides in the environment than do adults. Children have a greater surface to volume ratio than do adults; therefore, they receive a greater dose from the pesticides to which they are exposed. They metabolize toxicants slower than do adults, so the pesticide dose they receive remains with them longer.

The research reporting farmworker child pesticide exposure has considered only the concentrations of dialkylphosphate metabolites of organophosphorus (OP) pesticides (Barr et al. 2004), with one exception (Fenske et al. 2002). Analyses from Washington (Curl et al. 2002; Fenske et al. 2000a, 2000b; Koch et al. 2002; Loewenherz et al. 1997; Thompson et al. 2003), Oregon (Lambert et al. 2005), California (Bradman et al. 2005; Mills and Zahm 2001), the Rio Grande Valley of Texas (Shalat et al. 2003), and North Carolina (Arcury et al. 2005, 2006), indicate farmworker children are exposed to OP pesticides, and that the concentrations of dialkylphosphate metabolites in their urine are high. However, measurement of the
dialkylphosphate metabolites does not provide information on the specific OP pesticides to
which these children are exposed (Needham et al. 2005). Further, these analyses do not provide
information on the exposure of these children to the non-dialkylphosphate OP pesticides (e.g.,
acephate) or to other non-OP pesticides. While knowledge of the general levels of
dialkylphosphate OP pesticides in farmworker children is valuable, knowing specific pesticides
to which farmworker children are exposed is important because it will indicate the sources of this
exposure. For example, knowing that the metabolite specific to parathion is present in a large
percentage of children’s urine samples will direct efforts to identify, and then eliminate, the
sources of parathion exposure.

Only Fenske and colleagues (2002) have reported pesticide-specific metabolite levels in
urine samples collected from farmworker children. Focusing on the major metabolites of the OP
pesticides chlorpyrifos (TCPY) and parathion (PNP) among 75 children of farmworkers and
pesticide applicators, they found TCPY in urine samples from 18 (24%) of the children, and PNP
in urine samples from 5 (7%) of the children. PNP had no statistically significant predictors,
while the only statistically significant predictor of TCPY was living in a household that used an
OP pesticide in a garden.

Investigators have proposed a model of farmworker child pesticide exposure that includes
para-occupational, residential, and environmental factors (Fenske et al. 2005; Quandt et al.
2006). Para-occupational exposure results from contact with persons (e.g., parents) doing farm
work. Residential exposure results from contact with agricultural pesticides in the home, as well
as from the residential application of pesticides. Agricultural pesticides may be brought into the
dwelling by workers on clothing, boots, or containers, or directly applied to the dwelling.
Characteristics of the dwelling, such as amount of carpeting and general repair, can affect the
amount of pesticides that enter and accumulate there. Control of the dwelling and cleaning the
dwelling will reduce the amount of agricultural pesticides that accumulate. Environmental
exposure results from pesticides that are applied in the larger environment in which the child
lives, such as drift during application. Each form of exposure is moderated by safety behaviors
exercised by household residents and by characteristics of the child and parents. The
implementation of safety behaviors is more likely if parents have received safety training. Safety
behaviors include showering immediately after work, as well as storage and laundering of soiled
work clothes separate from the child’s clothing. Child characteristics, such as sex and age, will
modify exposure, as children of different ages and sex have different exposure behaviors.
Children of different ages metabolize pesticides differently. Parental characteristics that could
modify exposure behaviors include educational attainment and level of employment.

This analysis has two objectives. The first objective is to describe specific urinary
pesticide metabolite concentrations for young children living in farmworker households located
in eastern North Carolina and the number of metabolites detected for each child. The second
objective is to delineate the para-occupational, residential and environment risk factors
associated with the number of pesticide metabolites present. This analysis uses data collected
from 60 Latino children, 1 to 6 years of age, living in eastern North Carolina farmworker
households during the 2004 agricultural production season.

**Materials and Methods**

Data for this analysis were collected as part of *Casa y Campo*, a 4-year community-based
participatory environmental justice project in which environmental health scientists, health care
providers, and farmworker advocates collaborated to reduce pesticide exposure among
farmworkers and their families. *Casa y Campo* was implemented in a six county area of eastern
North Carolina, including Duplin, Harnett, Johnston, Sampson, Wake, and Wayne counties. For 2004, the North Carolina Employment Security Commission estimated that 21,614 migrant and seasonal farmworkers (not counting dependents) worked in these counties during peak harvest, accounting for one-quarter of the 86,040 migrant and seasonal farmworkers in the state. Agriculture in these six counties is diverse (Table 1). Most of the farms in these counties producing tobacco (1,329 farms), sweet potatoes (188 farms), and vegetables (435 farms) would employ migrant and seasonal farmworkers. All of these farms, as well as the large number of farms producing grains, soybeans, and cotton would use pesticides to which farmworkers and their families could be exposed. Results describing the dialkylphosphate urinary metabolites of OP pesticides for participants in this analysis have been published (Arcury et al. 2006).

Sample and Data Collection

Sampling, recruitment and data collection have been described in detail (Arcury et al. 2006). Briefly, from July through August, 2004, as part of a larger survey, 60 households with an adult resident who was employed in farm work and at least one resident child aged 1 to 6 years were recruited to participate. A site-based sampling approach was used to locate and recruit eligible participants (Arcury and Quandt 1999). All 60 eligible households agreed to participate. The mother of each child completed an interview questionnaire, and she was asked to collect a first morning urine void from the resident child between 1 and 6 years of age who was closest to age 5 years. Respondents were told they would receive a small gift, a bag of food, at the end of the study for completing the interview, and $10 for collecting the urine sample. All participants gave written informed consent. Study procedures were approved by the Wake Forest University School of Medicine and the Centers for Disease Control and Prevention’s (CDC) Institutional Review Boards.
The interviewer-administered questionnaires were completed in the respondents’ homes in their preferred language, which was Spanish in all cases. Data were collected on respondent and child characteristics, household characteristics, and dwelling quality. Interviews took approximately 25 minutes to complete. At the end of the interview, participants were asked to collect a first morning void from the selected child the next morning. Urine collection materials were left with participants. Each child’s urine sample was picked up from the home by a project staff member the morning the sample was collected and transported in a cooler with blue ice to our field laboratory, where it was frozen to -20ºC. In 7 instances, first morning voids were not collected, and the child participant provided a spot void.

Laboratory Analysis

The frozen urine samples were shipped overnight on dry ice to CDC in Atlanta, Georgia, for analysis. Samples were analyzed using a modification of the method of Olsson et al. (2004). Briefly, 2 mL urine samples were hydrolyzed by enzymes to liberate the glucuronide- or sulfate-bound conjugated metabolites. Hydrolysates were extracted using a mixed mode solid phase extraction cartridge. Concentrated extracts were analyzed using high performance liquid chromatography-tandem mass spectrometry. Two precursor/product ion pairs were analyzed per analyte, one for quantification and one for confirmation. Analyte concentrations were quantified using isotope dilution calibration. Approximately 10% of the samples tested were positive and negative quality control samples.

Measures

The outcome variables were based on pesticide metabolite concentrations obtained from each child’s urine sample. Metabolites included the OP insecticide metabolites TCPY, CMHC, IMPY, CIT, MDA, PNP, and DEAMPY, the pyrethroid insecticide metabolite 3PBA, DEET
repellent, and the herbicide metabolites 2,4,5-T, 2,4-D, ACE, ATZ and MET. Measures included the concentrations of each metabolite, and the total number of different pesticide metabolites detected in each child’s urine sample.

Measures were constructed for each of the domains in the model of pesticide exposure. Measures of para-occupational exposure included the variables whether mother was currently employed doing farm work, whether father was currently employed doing farm work, and number of farmworkers in the household (1, 2, 3 or more). Residential exposure measures included home ownership (rent, own, other), ease of cleaning (easy, difficult), number of bathrooms (1, 2, 3), and number of rooms with carpet (0, 1-2, 3 or more). Proximity to nearest agricultural field was the environmental exposure variable; it had the values of adjacent when agricultural fields directly abutted or were across the road from the property on which the respondent’s dwelling was located, and non-adjacent when agricultural fields did not abut the property on which the respondent’s dwelling was located.

Safety behavior measures included parental pesticide safety training (mother and father each coded no farm work, farm work and training, farm work and no training), whether any farmworker in the household regularly delayed showering after work for more than 15 minutes, whether any farmworker in the household regularly changed work clothes inside the dwelling, storage of soiled farm work clothes (everyone stores clothes outside, anyone stores clothes inside, anyone stores work clothes with other clothes), and whether everyone launders soiled farm clothes separately. Child and parental characteristics included child sex and age (1 or 2 years, 3 or 4 years, 5 or 6 years), mother’s education (6 or fewer years, 7 to 9 years, 10 or more years), and mother’s current employment (none, part time, full time).
Data Analysis

Medians and geometric means (when the number of detects was greater than 50%), unadjusted and adjusted for creatinine, were calculated for the individual 14 urinary pesticide metabolites. Concentrations below the analytic limits of detection (LOD) were substituted by the LOD divided by the square root of 2 (Homung and Reed 1990). To identify potential predictors for the number of detects, bivariate associations between the outcome and the predictor variables were examined through analysis of variance (ANOVA). Only the variables whose p-values were less than 0.20 were considered in further model selection process. Finally, a model with 3 variables (mother’s employment, child’s gender, and ownership) was prescribed by a forward selection procedure. Analyses were repeated with creatinine-adjusted measures to control for the 7 children who gave spot voids. The results did not differ, and the results for non-creatinine adjusted measures are presented. All analyses were performed using SAS 9.0 (SAS Institute, Inc.; Cary, NC).

Results

Exposure, Safety Behavior, and Personal Characteristics

The 60 children considered in this analysis varied in para-occupational, residential, and environmental sources of pesticide exposure, as well as safety behaviors and personal characteristics (Table 2). The majority of their mothers (68.3%) and fathers (56.7%) were employed as farmworkers at the time of data collection. Over two-thirds (36.7%) lived in households with two farmworkers, and 25.0% lived in households with three or more farmworkers. Most (53.3%) of the children lived in rented homes, while 33.4% lived in farmworker-owned homes and 13.3% lived in homes of other tenure. Over half (58.3%) lived in homes that their mothers described as easy to clean. Over half (53.3%) of these children lived in
dwellings with one bathroom. Most (91.7%) lived in dwellings with at least one carpeted room, and almost 70% lived in dwellings with three or more carpeted rooms. The dwellings in which over half (56.7%) of the children lived were not adjacent to agricultural fields.

Twenty-three of 41 mothers (38.3% of total sample) and 17 of the 34 fathers (28.3% of total sample) employed in farm work had not received pesticide safety training. Three-fifths of the children lived in households in which individuals employed in farm work did not shower immediately after work, and 83.3% lived in households in which individuals changed out of their soiled farm work clothes inside the dwelling. However, 15% lived in households in which farm work clothes were stored with other laundry, and 20% lived in households in which farm work clothes were laundered with other work clothes.

More of these children were girls (56.7%) than boys (43.3%). Most were 3 or 4 years of age (60.0%), with one-third 5 or 6 years of age, and 8.3% 1 or 2 years of age. Over half of their mothers had less than a secondary education. About one-quarter of their mothers did not work outside the home, with one-quarter working part-time and one-half working full-time.

**Pesticide Metabolite Levels**

The metabolites for 13 of the 14 pesticides were present in the urine samples of the 60 children (Table 3). Among these metabolites were those of 7 organophosphorus (OP) insecticides. PNP was the most common OP insecticide present; PNP was present in 90% of the samples, and had a geometric mean (GM) of 1.0 µg/L unadjusted for creatinine, and 1.25 µg/g adjusted for creatinine. TCPY was present in 83.3% of the samples, while IMPY was present in 55.0% of the samples, and MDA was present in 28.3% of the samples. The GM for TCPY was 1.92 µg/L unadjusted for creatinine, and 2.38 µg/g adjusted for creatinine; and the GM for IMPY was 0.56 µg/L unadjusted for creatinine, and 0.70 µg/g adjusted for creatinine. The median for
MDA was 0.21 µg/L unadjusted for creatinine, and 0.33 µg/g adjusted for creatinine. CMHC was detected in 7 (11.7%) of the samples, DEAMPY 3 (5.0%) of the samples, and CIT in 1 (1.7%) of the samples.

The pyrethroid insecticide metabolite 3PBA was present in 40.0% of the farmworker child urine samples. The metabolite for DEET repellent was present in 10.0% of the farmworker child urine samples. The herbicide 2,4,5-T was present in one of the samples. The herbicide 2,4-D was present in 41.7% of the samples, ACE was present in 21.7% of the samples, and ATZ was present in 6.7% of the samples.

**Pesticides Detected per Child: Number and Predictors**

The number of metabolites detected in the children’s urine samples varied from 0 to 7. One child (1.7%) had no detects, 5 children (8.3%) had one detect, 1 child (1.7%) had 2 detects, 16 children (26.7%) had 3 detects, 17 children (28.3%) had 4 detects, 8 children (13.3%) had 5 detects, 9 children (15.0%) had 6 detects, and 3 children (5.0%) had 7 detects.

Three of the predictors had statistically significant associations with the mean number of pesticides detected in the bivariate analysis (Table 4). Children residing in rented homes had 4.41 pesticides detected in their urine samples, compared to 3.65 detects among those living in an owned home and 3.00 detects among those living in other homes. Boys had 4.46 pesticides detects compared to 3.59 for girls. Finally, children of mothers working part-time had 5.06 detects, compared to 3.63 detects among children whose mothers worked full-time and 3.43 detects among children whose mothers did not work. In the multivariate analysis, mother’s employment remained significantly associated with number of detects, while the association of child sex trended toward significance and home ownership was not significant.
**Discussion**

Urine samples from most study children living in farmworker homes have a variety of pesticide metabolites. At least one specific pesticide metabolite was found in the urine samples for 59 of these 60 children; the urine samples for 88.3% of these children had three or more specific pesticide metabolites present. Analysis of the general dialkylphosphate OP pesticide metabolites for these same children found at least one of these six metabolites in every sample (Arcury et al. 2006).

The specific pesticide metabolites present indicate multiple sources and pathways of potential exposure among children in farmworker homes, and illustrate the length of time these pesticides remain in the environments of farmworker children. The pesticide metabolites present in these children’s urine indicate at least four pathways of exposure: (1) a para-occupational take-home pathway in which workers bring pesticides into their homes on their person or on their clothing; (2) an environmental pathway in which pesticides applied to nearby fields drift into the residential environment; (3) a residential pathway of pesticides applied in the home; and (4) a residual pathway in which pesticides deposited inside the home from any of the pathways at an earlier time remain active. Earlier research in North Carolina showed that farmworkers had little knowledge of the pesticides used where they work, and that they did not have information about pesticides applied to dwellings that were rented or grower-provided (Arcury et al. 2001; Early et al. 2006). Drift has been shown to result in the dispersal of agricultural pesticides to dwellings on surrounding land (Fenske et al. 2000b; Ward et al. 2006; Weppner et al. 2006). Finally, agricultural communities have generally higher levels of environmental pesticides compared to non-agricultural communities, with residues from discontinued pesticides found in contemporary rural environments and homes (Lu et al. 2004; Quandt et al. 2004; Wolz et al.)
However, the possible sources and pathways of exposure for several pesticide metabolites found in the samples from these children remain unclear.

In discussing the sources and pathways of potential exposure it is important to remember that the urine samples used in this analysis were collected in June and July, 2004. Exposure of these children to pesticides in Mexico should be discounted as the source of the metabolites; these are the children of seasonal rather than migrant farmworkers, who have been settled in North Carolina for several years. Many of these children were born in the U.S., and have limited direct or indirect contact with sources of exposure from Mexico.

Chlorpyrifos and diazinon have been used for residential as well as agricultural applications. However, chlorpyrifos was banned for residential use by the end of 2001, while diazinon was banned for residential use at the end of 2004. Fifty of the 60 children had TCPY, the chlorpyrifos metabolite, in their urine. This would indicate that chlorpyrifos exposure for these children did not result from recent residential application, but could result from the take-home and drift pathways, as well as from residual deposition. The diazinon metabolite IMPY was present in 33 of the 60 samples. Because it was still available for residential use at the time these data were collected, it is possible that diazinon exposure could result from any of the pathways.

Malathion has outdoor residential use, while parathion has no residential use. Yet 54 of the 60 children had the parathion metabolite PNP and 17 had the malathion metabolite MDA. Farmworkers may be exposed to malathion at work (e.g., picking peppers), but it is unlikely that they would work in fields to which parathion has been applied (e.g., cotton). Therefore, exposure resulting in the presence of PNP is most likely the result of the drift and residual deposition pathways, while MDA could result from the take-home, drift, or residual pathways.
Pyrethroid insecticides are widely used for residential and agricultural applications. The 24 children with 3PBA, the pyrethorid metabolite, in their urine sample could be exposed through any of the pathways. The herbicides 2,4-D and acetochlor are widely used for residential and agricultural applications, and their metabolites could result from exposure through any of the pathways.

Atrazine is used on corn (85% of all usage), sorghum (10%), and sugar cane (3%). Three-quarters of all corn has atrazine applied. Corn and sorghum are grown in North Carolina. However, field corn and sorghum are machine cultivated and harvested. Sweet corn is hand picked, but few farmworkers are employed for this task. The most plausible explanation for 4 farmworker children having the metabolite for atrazine in their urine can be limited to drift or contaminated water supplies.

The sources and pathways of exposure among farmworker children to the OP pesticides coumaphas, pirimiphos methyl, and isazophos, and to the herbicide 2,4,5-T are not apparent. Coumaphas is used to control insects on livestock, and farmworkers in North Carolina seldom work with livestock. However, CMHC, the metabolite of coumaphas, was detected in samples from 7 of the children. Pirimiphos methyl is used post-harvest to treat stored corn and sorghum grain, activities with limited farmworker participation. Yet, DEAMPY, the metabolite of pirimiphos methyl, was detected in samples from 3 children. Isazophos was used to treat lawns and turf; isazophos was cancelled in North Carolina at the end of 1998, and cancelled by the U.S. Environmental Protection Agency in mid-1999. Yet, CIT, the metabolite of isazophos was found in one child. Finally, the herbicide 2,4,5-T has been banned for use in the U.S. since 1986, years before most Latino farmworkers had immigrated to North Carolina. The metabolite for 2,4,5-T was detected in the sample of one child. The metabolites for isazophos and 2,4,5-T were found
in samples from different children.

Few data exist with which our results can be compared. The 1999-2000 NHANES data provide results for the proportion of detects and levels detected for the OP pesticide metabolites TCPY, IMPY, MDA, and PNP (Barr et al. 2005). Caution must be taken in the comparison of our results with those reported for the NHANES sample. Development is a major factor affecting pesticide metabolism; and our participants were aged 1-6 years, while the youngest age group included in the NHANES sample was 6-11 years (Table 5). For TCPY, the percent of children with pesticide metabolite detected in the entire 1999-2000 NHANES sample (91%), among the Mexican-American sample (87%), and among children aged 6 to 11 years (97%) was somewhat greater than for the children in our sample (83%), with the GMs, 50th percentile, and 95th percentile (creatinine unadjusted and adjusted) being lower for the entire and Mexican American samples, but higher among children aged 6 to 11 years. The percent detects for IMPY and PNP was higher for our sample (55% and 90%, respectively) compared to the NHANES samples (total sample, 29%, 22%; Mexican-American sample, 24%, 34%; children aged 6 to 11 years sample, 26%, 26%). The 95th percentile of IMPY for our sample (3.94 µg/L unadjusted for creatinine, 3.01 µg/g adjusted for creatinine) approximates those for the NHANES samples. The 95th percentile for PNP is greater among the children in our sample (6.32 µg/L unadjusted for creatinine, 7.99 µg/g adjusted for creatinine) than for the entire NHANES sample (5.0, 4.2) and among children aged 6 to 11 years (4.2, 4.2). However, the 95th percentile is much higher in the NHANES Mexican-American sample (21, 17). The percent detects for MDA for the children in our sample (28%) is lower than the percent detects in the entire 1999-2000 NHANES sample 52%), among the Mexican-American Sample (61%), and among children aged 6 to 11 years (56%). However, the 95th percentile (6.87 µg/L unadjusted for creatinine, 8.64 µg/g adjusted for
creatinine) for the children in our sample is greater than that for the entire NHANES sample (1.6, 1.8), the NHANES Mexican-American sample (1.6, 1.7), and among children aged 6 to 11 years (2.8, 3.7). No comparison data are available for the pyrethroid insecticide metabolite 3PBA, DEET repellent, or any of the herbicide metabolites.

Fenske and colleagues (2002) report on the presence and levels of two OP pesticide metabolites for children residing in Washington State agricultural households. Not all of the children lived in farmworker households. They found TCPY in urine samples from 24% (18 of 75) children, with level of 6.0 µg/L (SD = 17) among children residing within 200 feet of an agricultural field and 1.3 µg/L (SD =4.9) for those living more than 200 feet from a field. This metabolite was present in 83.3% of our samples, and had a GM of 1.92 µg/L. They reported PNP in urine samples for 7% (5 of 75) children. The comparable mean PNP levels were 33 µg/L (SD=210) and 0. This metabolite was present in 90.0% of our samples, and had a GM of 1.00 µg/L. These differences reflect regional differences in agricultural uses.

Earlier studies have not considered the number of different pesticide metabolites present in individual children. Combining doses from several pesticides in a single child may result in interactions beyond the effects of a single toxicant (Eskenazi et al. 1999). Among these 60 children, 54 had 2 or more pesticide metabolites, with 9 children having 6 pesticide metabolites and 3 children having 7 pesticide metabolites. There was no discernable pattern in the combinations of pesticide metabolites present in the samples. Boys versus girls, and those living in rented housing versus owned or other housing had a greater number of pesticide metabolites in bivariate analysis. This may reflect gender differences in behavior. More importantly, it reflects how control of a dwelling may decrease exposure. Children whose mothers worked part-time versus full-time or who did not work had a greater number of pesticide metabolites in the
bivariate and multivariate analysis. The interpretation of this association is not clear. Further analysis was conducted to examine whether mothers working part-time were more likely to be employed doing farmwork; this was not the case. It is possible that mothers working part-time are less able than unemployed women to attend to household hygiene, while mothers working full-time could count on more help in home care. Research is needed that has greater precision in measuring of pesticide exposure risk factors (Arcury et al. 2006; Quandt et al. 2006).

The results of this study should be considered in light of its limitations. The cross-sectional design limits analysis to association rather than causation. The lack of environmental pesticide samples does not allow for consideration of the location of exposure. Data on residential pesticide application were not collected. The metabolites present in the urine may also include exposure to the less toxic environmental degradates of the target pesticides. The determination of whether a metabolite was detected in a child’s urine sample, and therefore the total number of metabolites detected in a child, maybe limited by the LOD of current analytical procedures; therefore, the number of metabolites detected is a conservative estimate. Finally, the measurement of risk factors may lack precision. This study also has several strengths. It is based on a relatively large sample of children. It is one of the few studies of farmworker child exposure on the East Coast. Finally, it is one of only studies to examine a large number of specific pesticide metabolites for farmworker children.

Several implications for environmental and occupational health practice and research can be drawn from our results. Pesticide exposure pathways for farmworkers and their children are multiple and complex. Pathways include take-home, drift, residential application, and residues. The sources of exposure to several of the pesticide metabolites found in the urine samples are not clear. Farmworkers must be educated about the take-home pathway. Current United States
Environmental Protection Agency Worker Protection Standard training (U.S. EPA 1992), when workers receive it (Arcury et al. 2001; U.S. GAO 2000), provides little information about pesticides that can be taken home on the clothes, boots and skin of workers, and about how to reduce or eliminate the amount taken home. Regulations on pesticide application to reduce drift must be reviewed. Farmworker and all rural families must be educated about drift and how to reduce exposure. Farmworker housing regulations must be improved to reduce need for residential pesticide application (Early et al. 2006; Housing Assistance Council 2001). Farmworkers need to be educated about alternatives to using pesticides, such as residential integrated pest management. Farmworker and other rural homes need to be deep cleaned to remove pesticides from indoor environments. Procedures for this deep cleaning need to be developed and tested (McCauley et al. 2006a).

Research is needed that includes more precise measures of exposure pathways. Laboratory techniques measuring pesticide metabolites in urine and other matrices, as well as environmental monitoring have improved substantially (Barr et al. 2006; Hoppin et al. 2006). However, measurement of exposure pathways in epidemiological research has not improved greatly. For example, asking about recent employment as a farmworker needs to be changed to the amount of farm work conducted in the three days before samples for biomonitoring are collected.

Research on the health effects of concurrent exposure to multiple individual pesticides and classes of pesticides is also needed. Current knowledge of health effects of pesticides in general is limited, but documentation is beginning (McCauley et al. 2006b). Examining health effects in light of the interactions of several different pesticides will provide greater insight into the actual risks to those exposed.
References

Arcury TA, Grzywacz JG, Davis SW, Barr DB, Quandt SA. 2006. Organophosphorus pesticide urinary metabolite levels of children in farmworker households in eastern North Carolina. Am J Ind Med 49:751-760.

Arcury TA, Quandt SA. 1999. Participant recruitment for qualitative research: a site-based approach to community research in complex societies. Hum Organ 58:128-133.

Arcury TA, Quandt SA, Cravey AJ, Elmore RC, Russell GB. 2001. Farmworker reports of pesticide safety and sanitation in the work environment. Am J Ind Med 39:487-498.

Arcury TA, Quandt SA, Rao P, Doran A, Snively BM, Davis S, et al. 2005. Organophosphate pesticide exposure in farmworker family members in western North Carolina and Virginia: case comparisons. Hum Organ 64:40-51.

Barr DB, Allen R, Olsson AO, Bravo R, Caltabiano LM, Montesano A, et al. 2005. Concentrations of selective metabolites of organophosphorus pesticides in the United States population. Environ Res 99:314-326.

Barr DB, Bravo R, Weerasekera G, Caltabiano LM, Whitehead RD Jr, Olsson AO, et al. 2004. Concentrations of dialkylphosphate metabolites of organophosphorus pesticides in the US population. Environ Health Perspect 112:186-200.

Barr DB, Thomas D, Curwin B, Landsittel D, Raymer J, Lu C, et al. 2006. Biomonitoring of exposure in farmworker studies. Environ Health Perspect 114:936-942.

Bradman A, Eskenazi B, Barr DB, Bravo R, Castorina R, Chevrier J, et al. 2005. Organophosphate urinary metabolite levels during pregnancy and after delivery in women living in an agricultural community. Environ Health Perspect 113:1802-1807.

Coronado GD, Thompson B, Strong L, Griffith WC, Islas I. 2004. Agricultural task and exposure
to organophosphate pesticides among farmworkers. Environ Health Perspect 112:142-147.

Curl CL, Fenske RA, Kissel JC, Shirai JH, Moate TF, Griffith W, et al. 2002. Evaluation of take-home organophosphorus pesticide exposure among agricultural workers and their children. Environ Health Perspect 110:A787-A792.

Early J, Davis SW, Quandt SA, Rao P, Snively BM, Arcury TA. 2006. Housing characteristics of farmworker families in North Carolina. J Immigr Minor Health 8:173-184.

Eskenazi B, Bradman A, Castorina R. 1999. Exposures of children to organophosphate pesticides and their potential adverse health effects. Environ Health Perspect 107(Suppl 3):409-419.

Faustman EM, Silbernagel SM, Fenske RA, Burbacher TM, Ponce RA. 2000. Mechanisms underlying children's susceptibility to environmental toxicants. Environ Health Perspect 108(Suppl 1):13-21.

Fenske RA, Kissel JC, Lu C, Kalman DA, Simcox NJ, Allen EH, et al. 2000a. Biologically based pesticide dose estimates for children in an agricultural community. Environ Health Perspect 108:515-520.

Fenske RA, Lu C, Barr D, Needham L. 2002. Children’s exposure to chlorpyrifos and parathion in an agricultural community in central Washington State. Environ Health Perspect 110:549-553.

Fenske RA, Lu C, Curl CL, Shirai JH, Kissel JC. 2005. Biologic monitoring to characterize organophosphorus pesticide exposure among children and workers: An analysis of recent studies in Washington State. Environ Health Perspect 113:1651-1657.

Fenske RA, Lu C, Simcox NJ, Loewenherz C, Touchstone J, Moate TF, et al. 2000b. Strategies for assessing children’s organophosphorus pesticide exposures in agricultural communities. J Expo Anal Environ Epidemiol 10:662-671.
Homung RW, Reed DL. 1990. Estimation of average concentration in the presence of nondetectable values. Appl Occup Environ Hyg 5:46-51.

Hoppin JA, Adgate JL, Eberhart M, Nishioka M, Ryan PB. 2006. Environmental exposure assessment of pesticides in farmworker homes. Environ Health Perspect 114:929-935.

Housing Assistance Council. 2001 No refuge from the fields: Findings from a survey of farmworker housing conditions in the United States. Washington, DC: Housing Assistance Council.

Koch D, Lu C, Fisker-Andersen J, Jolley L, Fenske RA. 2002. Temporal association of children's pesticide exposure and agricultural spraying: Report of a longitudinal biological monitoring study. Environ Health Perspect 110:829-833.

Lambert WE, Lasarev M, Muniz J, Scherer J, Rothlein J, Santana J, et al. 2005. Variation in organophosphate pesticide metabolites in urine of children living in agricultural communities. Environ Health Perspect 113:504-508.

Loewenherz C, Fenske RA, Simcox NJ, Bellamy G, Kalman D. 1997. Biological monitoring of organophosphorus pesticide exposure among children of agricultural workers in central Washington State. Environ Health Perspect 105:1344-1353.

Lu C, Kedan G, Fisker-Andersen J, Kissel JC, Fenske RA. 2004. Multipathway organophosphorus pesticide exposures of preschool children living in agricultural and nonagricultural communities. Environ Res 96:283-289.

McCauley LA, Anger WK, Keifer M, Langley R, Robson MG, Rohlman D. 2006b. Studying health outcomes in farmworker populations exposed to pesticides. Environ Health Perspect 114:953-960.

McCauley LA, Travers R, Lasarev M, Muniz J, Nailon R. 2006a. Effectiveness of cleaning
procedures in removing pesticides from home environments. J Agromedicine 11:81-88.
Mills PK, Zahm S. 2001. Organophosphate pesticide residues in urine of farmworkers and their children in Fresno County, CA. Am J Ind Med 40:571-577.

Needham LL, Barr DB, Caudill SP, Pirkle JL, Turner WE, Osterloh J, et al. 2005. Concentrations of environmental chemicals associated with neurodevelopmental effects in U.S. population. Neurotoxicology 26:531-545.

Olsson AO, Baker SE, Nguyen JV, Romanoff LC, Udunka SO, Walker RD, et al. 2004. A liquid chromatography--tandem mass spectrometry multiresidue method for quantification of specific metabolites of organophosphorus pesticides, synthetic pyrethroids, selected herbicides, and DEET in human urine. Anal Chem 76:2453-2461.

Quandt SA, Arcury TA, Rao P, Snively BM, Camann DE, Doran AM, et al. 2004. Agricultural and residential pesticides in wipe samples from farmworker family residences in North Carolina. Environ Health Perspect 112:382-387.

Quandt SA, Hernández-Valero MA, Grzywacz JG, Hovey JD, Gonzales M, Arcury TA. 2006. Workplace, household, and personal predictors of pesticide exposure for farmworkers. Environ Health Perspect 114:943-952.

Reigart JR, Roberts JR. 1999. Recognition and Management of Pesticide Poisonings. 5th ed. Washington DC: Environmental Protection Agency.

Shalat SL, Donneley KC, Freeman NC, Calvin JA, Ramesh S, Jimenez M, et al. 2003. Nondietary ingestion of pesticides by children in an agricultural community on the US/Mexico border: preliminary results. J Expo Anal Environ Epidemiol 13:42-50.

Strong LL, Thompson B, Coronado GD, Griffith WC, Vigoren EM, Islas I. 2004. Health symptoms and exposure to organophosphate pesticides in farmworkers. Am J Ind Med
46:599-606.

Thompson B, Coronado GD, Grossman JE, Puschel K, Solomon CC, Curl CL, et al. 2003. Pesticide take-home pathways among children of agricultural workers: Study design, methods and baseline findings. J Occup Environ Med 45:42-53.

USDA (U.S. Department of Agriculture). 2007. Quik Stats: Agricultural Statistics Data Base. Available: http://www.nass.usda.gov/Census/ [accessed 11 March 2007].

U.S. EPA. 1992. Worker Protection Standard. Federal Register Title 40 CFR, Part 70. Washington, DC: U.S. Environmental Protection Agency.

U.S. GAO. 2000. Pesticides: Improvements needed to ensure the safety of farmworkers and their children. GAO/RCED-00-40. Washington, DC: U.S. Government Accountability Office.

Ward MH, Lubin J, Giglierano J, Colt JS, Wolter C, Bekiroglu N, et al. 2006. Proximity to crops and residential exposure to agricultural herbicides in Iowa. Environ Health Perspect 114:893-897.

Weiss B, Amler S, Amler RW. 2004. Pesticides. Pediatrics 113:1030-1036.

Weppner S, Elgethun K, Lu C, Hebert V, Yost MG, Fenske RA. 2006. The Washington aerial spray drift study: children's exposure to methamidophos in an agricultural community following fixed-wing aircraft applications. J Expo Sci Environ Epidemiol 16:387-396.

Wolz S, Fenske RA, Simcox NJ, Palcisko G, Kissel JC. 2003. Residential arsenic and lead levels in an agricultural community with a history of lead arsenate use. Environ Res 93:293-300.
| County  | Number of Farms | Acres in Farmland | Tobacco | Sweet Potatoes | Vegetables | Corn for Grain | Wheat for Grain | Soybeans | Cotton |
|---------|-----------------|-------------------|---------|----------------|------------|----------------|----------------|----------|--------|
| Duplin  | 1,190           | 234,658           | 213     | 9              | 74         | 341            | 137            | 401      | 58     |
| Harnett | 740             | 114,361           | 144     | 19             | 55         | 157            | 44             | 173      | 38     |
| Johnston| 1,144           | 194,211           | 360     | 91             | 76         | 152            | 108            | 431      | 56     |
| Sampson | 1,178           | 298,483           | 243     | 44             | 135        | 279            | 162            | 431      | 126    |
| Wake    | 846             | 92,803            | 161     | 18             | 51         | 22             | 54             | 102      | 1      |
| Wayne   | 722             | 171,449           | 208     | 7              | 44         | 187            | 134            | 296      | 80     |

*aUSDA 2007.*
Table 2: Exposure Characteristics of Children Aged 1 to 6 Years Living in Farmworker Families, Eastern North Carolina, 2004.

| Exposure Characteristics                                      | N   | %    |
|---------------------------------------------------------------|-----|------|
| **Para-Occupational Exposure**                                |     |      |
| Mother does farm work                                         | 41  | 68.3 |
| Father does farm work                                         | 34  | 56.7 |
| Number of farmworkers in house                                |     |      |
| 1                                                             | 23  | 38.3 |
| 2                                                             | 22  | 36.7 |
| 3 or more                                                     | 15  | 25.0 |
| **Residential Exposure**                                      |     |      |
| Home ownership                                                |     |      |
| Own                                                           | 20  | 33.3 |
| Rent                                                          | 32  | 53.3 |
| Other                                                         | 8   | 13.3 |
| Ease of cleaning                                              |     |      |
| Easy                                                          | 35  | 58.3 |
| Difficult                                                     | 25  | 41.7 |
| Number of bathrooms                                           |     |      |
| 1                                                             | 32  | 53.3 |
| 2 or 3                                                        | 28  | 46.7 |
| Number of rooms with carpet                                   |     |      |
| 0                                                             | 5   | 8.3  |
| 1 or 2                                                        | 14  | 23.4 |
| 3 or more                                                     | 41  | 68.3 |
| **Environmental Exposure**                                    |     |      |
| Proximity to nearest agricultural field                       |     |      |
| Adjacent                                                      | 26  | 43.3 |
| Nonadjacent                                                   | 34  | 56.7 |
| **Safety Behaviors**                                          |     |      |
| Mother’s pesticide safety training                            |     |      |
| No farm work                                                  | 19  | 31.7 |
| Farm work and training                                        | 18  | 30.0 |
| Farm work and no training                                     | 23  | 38.3 |
| Father’s pesticide safety training                            |     |      |
| No farm work                                                  | 26  | 43.3 |
| Farm work and training                                        | 11  | 18.3 |
| Farm work and no training                                     | 17  | 28.3 |
| Anyone in house delays shower after farm work                 | 36  | 60.0 |
| Anyone in house changes farm clothes inside                   | 50  | 83.3 |
| Farm work clothes storage                                     |     |      |
| Everyone stores clothes outside                               | 11  | 18.3 |
| Anyone stores clothes inside                                  | 40  | 66.7 |
| Anyone stores clothes with other clothes                      | 9   | 15.0 |
| Everyone launders farm clothes separately                     | 48  | 80.0 |
| Child and Mother Characteristics |       |       |
|----------------------------------|-------|-------|
| **Child Sex**                    |       |       |
| Female                           | 34    | 56.7  |
| Male                             | 26    | 43.3  |
| **Child Age**                    |       |       |
| 1 or 2 years                     | 5     | 8.3   |
| 3 or 4 years                     | 36    | 60.0  |
| 5 or 6 years                     | 19    | 31.7  |
| **Mother’s Education**           |       |       |
| 1 to 6 years                     | 34    | 56.7  |
| 7 to 9 years                     | 15    | 25.0  |
| 10 or more years                 | 11    | 18.3  |
| **Mother’s Employment**          |       |       |
| None                             | 14    | 23.3  |
| Part time                        | 16    | 26.7  |
| Full time                        | 30    | 50.0  |
Table 3: Pesticide Urinary Metabolites for Children Age 1 to 6 Years (n=60) in Eastern North Carolina Farmworker Households, 2004: Limit of Detection, Number of Detects, Median and Geometric Mean, and Creatinine Adjusted Median and Geometric Mean.

| Pesticide Urinary Metabolites | Limit of Detection (ng/mL) | Detectors | Not Adjusted for Creatinine (µg/L) | Creatinine Adjusted (µg/g creatinine) |
|-------------------------------|---------------------------|-----------|-----------------------------------|--------------------------------------|
|                              |                           | N         | Median   | Geometric Mean                      | Median      | Geometric Mean |
| TCPY (Chlorpyrifos)          | 0.2                       | 50        | 83.3     | 2.47                                | 1.92        | 3.38           | 2.38           |
| CMHC (Coumaphos)             | 0.2                       | 7         | 11.7     | 0.14                                | a           | 0.17           | a              |
| IMPY (Diazinon)              | 0.7                       | 33        | 55.0     | 0.49                                | 0.56        | 0.60           | 0.70           |
| CIT (Isazophos)              | 1.5                       | 1         | 1.7      | 1.06                                | a           | 1.19           | a              |
| MDA (Malathion)              | 0.3                       | 17        | 28.3     | 0.21                                | a           | 0.33           | a              |
| PNP (Parathion)              | 0.1                       | 54        | 90.0     | 1.55                                | 1.00        | 1.53           | 1.25           |
| DEAMPY (Pirimiphos methyl)   | 0.2                       | 3         | 5.0      | 0.14                                | a           | 0.17           | a              |
| Pyrethroid Insecticides      |                           |           |          |                                     |             |                |                |
| 3PBA                         | 0.1                       | 24        | 40.0     | 0.07                                | a           | 0.15           | a              |
| DEET Repellent               | 0.1                       | 6         | 10.0     | 0.07                                | a           | 0.08           | a              |
| Herbicides                   |                           |           |          |                                     |             |                |                |
| 2,4,5-T                      | 0.1                       | 1         | 1.7      | 0.07                                | a           | 0.08           | a              |
| 2,4-D                        | 0.2                       | 25        | 41.7     | 0.14                                | a           | 0.23           | a              |
| ACE (Acetochlor)             | 0.1                       | 13        | 21.7     | 0.07                                | a           | 0.09           | a              |
| ATZ (Atrazine)               | 0.3                       | 4         | 6.7      | 0.09                                | a           | 0.10           | a              |
| MET (Metolachlor)            | 0.2                       | 0         | 0.0      | --                                  | --          | --             | --             |

*aThe small number of detects (less than 50%) make the geometric mean an unreliable measure for this metabolite.
Table 4: Predictors with a Significant Relationship to Number of Pesticide Urinary Metabolite Detects, Children Age 1 to 6 Years (n=60) in Eastern North Carolina Farmworker Households, 2004.

| Variable               | Number of Detects | Significance | Bivariate Analysis \(^a\) | Multivariate Analysis \(^a\) |
|------------------------|-------------------|--------------|-----------------------------|------------------------------|
|                       | Mean   | SE     |                             |                              |
| Home Ownership         |        |        |                             |                              |
| Own                    | 3.65   | 0.35   | 0.045                       | ---                          |
| Rent                   | 4.41   | 0.27   |                             |                              |
| Other                  | 3.00   | 0.54   |                             |                              |
| Child Sex              |        |        |                             |                              |
| Male                   | 4.46   | 0.31   | 0.035                       | 0.078                        |
| Female                 | 3.59   | 0.27   |                             |                              |
| Mother’s Employment    |        |        |                             |                              |
| None                   | 3.43   | 0.39   | 0.004                       | 0.008                        |
| Part-time              | 5.06   | 0.37   |                             |                              |
| Full-time              | 3.63   | 0.27   |                             |                              |

\(^a\)Other variables considered in the bivariate analysis and included in the multivariate analysis were mother does farm work, father does farm work, number of farmworkers in house, ease of cleaning house, number of bathrooms, number of rooms with carpet, proximity to nearest agricultural field, mother’s pesticide safety training, father’s pesticide safety training, anyone in house delays shower after farm work, anyone in house changes farm clothes inside, farm work clothes storage, everyone launders farm clothes separately, child age, and mother’s education.
Table 5: Comparison of Percent Detects, Creatinine Unadjusted and Adjusted Geometric Mean, 50th Percentile, and 95th Percentile, for Children Aged 1 – 6 Years (n=60) in Eastern North Carolina Farmworker Households (Age 1-6), 2004, and 1999-2000 NHANES Data for Total, Mexican-American, and Children Aged 6 – 11 Years.

| Metabolites and Samples | Creatinine Unadjusted | Creatinine Adjusted |
|-------------------------|-----------------------|---------------------|
|                         | Percent Detects | Geometric Mean | 50th Percentile | 95th Percentile | Geometric Mean | 50th Percentile | 95th Percentile |
| TCPY                    | Eastern North Carolina Farmworker Children | 83.3 | 1.92 | 2.47 | 16.91 | 2.38 | 3.4 | 17.46 |
|                         | Total | 91 | 1.77 | 1.7 | 9.9 | 1.58 | 1.47 | 8.4 |
|                         | Mexican-American | 87 | 1.61 | 1.67 | 7.4 | 1.46 | 1.44 | 5.8 |
|                         | Children aged 6 to 11 | 97 | 2.88 | 2.7 | 16 | 3.11 | 3.2 | 14 |
| IMPY                    | Eastern North Carolina Farmworker Children | 55.0 | 0.56 | 0.49 | 3.94 | 0.70 | 0.60 | 3.01 |
|                         | Total | 29 | -- | -- | 3.7 | -- | -- | 3.4 |
|                         | Mexican-Americans | 24 | -- | -- | 4.2 | -- | -- | 3.9 |
|                         | Children aged 6 to 11 | 26 | -- | -- | 3.6 | -- | -- | 5.1 |
| MDA                     | Eastern North Carolina Farmworker Children | 28.3 | -- | 0.21 | 6.87 | 0.57 | 0.34 | 8.64 |
|                         | Total | 52 | -- | <LOD | 1.6 | -- | <LOD | 1.8 |
|                         | Mexican-Americans | 61 | -- | <LOD | 1.6 | -- | <LOD | 1.7 |
|                         | Children aged 6 to 11 | 56 | -- | 0.49 | 2.8 | -- | 0.44 | 3.7 |
| PNP                     | Eastern North Carolina Farmworker Children | 90.0 | 1.00 | 1.55 | 6.32 | 1.25 | 1.52 | 7.99 |
|                         | Total | 22 | -- | -- | 5 | -- | -- | 4.2 |
|                         | Mexican-Americans | 34 | -- | -- | 21 | -- | -- | 17 |
|                         | Children aged 6 to 11 | 26 | -- | -- | 4.2 | -- | -- | 4.2 |

aBarr et al. 2005.