Functional Plant Types Drive Plant Interactions in a Mediterranean Mountain Range

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Shrubs have positive (facilitation) and negative (competition) effects on understory plants, the net interaction effect being modulated by abiotic conditions. Overall shrubs influence to great extent the structure of plant communities where they have significant presence. Interactions in a plant community are quite diverse but little is known about their variability and effects at community level. Here we checked the effects of co-occurring shrub species from different functional types on a focal understory species, determining mechanisms driving interaction outcome, and tested whether effects measured on the focal species were a proxy for effects measured at the community level. Growth, physiological, and reproductive traits of Euphorbia nicaeensis, our focal species, were recorded on individuals growing in association with four dominant shrub species and in adjacent open areas. We also recorded community composition and environmental conditions in each microhabitat. Shrubs provided environmental conditions for plant growth, which contrasted with open areas, including moister soil, greater N content, higher air temperatures, and lower radiation. Shrub-associated individuals showed lower reproductive effort and greater allocation to growth, while most physiological traits remained unaffected. Euphorbia individuals were bigger and had more leaf N under N-fixing than under non-fixing species. Soil moisture was also higher under N-fixing shrubs; therefore soil conditions in the understory may counter reduced light conditions. There was a significant effect of species identity and functional types in the outcome of plant interactions with consistent effects at individual and community levels. The contrasting allocation strategies to reproduction and growth in Euphorbia plants, either associated or not with shrubs, showed high phenotypic plasticity and evidence its ability to cope with contrasting environmental conditions.

Keywords: Biomass allocation, competition, facilitation, functional traits, plant interaction balance, phenotypic plasticity, Sierra Nevada Mountains, water availability

INTRODUCTION

Plant interactions modulate the structure of plant communities and shape species distribution (Callaway, 2007; Butterfield et al., 2010; Cavieres et al., 2014). Some plant species facilitate establishment and growth of other species through amelioration of physical stress (Moro et al., 1997; le Roux and McGeoch, 2010) or resource supply (Pugnaire et al., 1996; Prieto et al., 2012)
while competition may counter facilitation effects determining net interaction outcomes (Tielbörger and Kadmon, 2000; Pugnaire and Luque, 2001; Armas et al., 2011). Indeed, while positive interactions (i.e., facilitation) enhance plant growth, reproduction, and survival of understory species, eventually expanding their distribution range (le Roux et al., 2012), negative interactions (i.e., competition, interference) limit growth and fitness of other species and may even completely exclude them from suitable habitats (Choler et al., 2001; le Roux et al., 2013). Overall, there is now ample evidence of plant interactions enhancing coexistence in plant communities at local (Pugnaire et al., 1996; Holzapfel and Mahall, 1999; Choler et al., 2001; Maestre et al., 2003; le Roux and McGeoch, 2010; Schöb et al., 2012; Tirado et al., 2015) and global scales (Callaway et al., 2002; Kikvidze et al., 2005; Butterfield et al., 2013; He et al., 2013; Cavieres et al., 2014). Yet there is less evidence on the intensity of plant interactions within a plant community in a given environment (Pugnaire et al., 2004; Callaway, 2007).

Most studies on plant interactions focused on the effects of a particular species on one (Callaway et al., 1996; Pugnaire et al., 1996; Shultz et al., 2007) or several understory species (Rodriguez-Echeverría and Pérez-Fernández, 2003; Gómez-Aparicio et al., 2004; Padilla and Pugnaire, 2009; Pistón et al., 2015). In plant communities, where usually several dominant species co-exist, there is little information on how interaction outcome may vary across species (Pugnaire et al., 2004; but see Liancourt et al., 2005; Schöb et al., 2012; Pistón et al., 2016). In addition, there is a growing body of evidence suggesting that certain species are more likely to act as facilitators than others (Callaway, 2007; Paterno et al., 2016), and that these differences may be result from e.g., specific functional traits related to plant phylogeny (Garbin et al., 2014) or differential trait effect on surrounding abiotic conditions (Schöb et al., 2013a). Several mechanisms can generate species-specific and plant functional type specific facilitative relationships linked to the variety of ways facilitator species influence their environment and resources, e.g., increasing soil water or nutrient content (Callaway, 1998). For instance, several studies have found that neighboring plants growing close to leguminous species can benefit from the additional N supply to the soil, resulting in greater N concentration in their leaves (Pugnaire et al., 1996; Temperton et al., 2007). Thus, species belonging to different functional types (e.g., N-fixers and non-fixers) may differently affect performance of understory species, resulting in a range of interactions that could go from competition to facilitation within the same community (Pugnaire et al., 2004). Moreover, whether such effects are consistent at the species and community levels remains unclear (Soliveres et al., 2015).

In stressful environments, such as dry mountains, shrubs can modify the environment under their canopy by improving microclimatic conditions (Gómez-Aparicio et al., 2004; Schöb et al., 2013a) or increase availability of soil resources (Verdú and García-Fayos, 2003; Gazol and Camarero, 2012). There is also evidence that shrubs may improve conditions for the establishment and growth of other species under their canopy despite strong resource competition (Holmgren et al., 1997; Moro et al., 1997; Schöb et al., 2014a,b; Tirado et al., 2015). These mechanisms operate simultaneously (Pugnaire and Luque, 2001), exerting both positive and negative effects (e.g., improving water availability while reducing light under the canopy) that would ultimately shape plant interaction outcome. However, we have less evidence about the factors that are more important in driving plant interactions in dry mountain systems. In the southern slopes of the Sierra Nevada Mountains in Spain, at elevations below 2500 m, precipitation is low and temperature and radiation high, leading to rather demanding conditions for plants (Callaway et al., 2002; Sánchez-Marañón et al., 2002; Schöb et al., 2013a). This system provides an opportunity to assess the contribution of different factors (e.g., water, nutrients, light or temperature) to plant interaction outcome. The co-occurrence of several shrub species with contrasting functional types (e.g., N fixers and non-fixers) makes this system suitable to address species-specific effects on understory species.

Our aim was to assess the effects of shrubs of contrasting functional type on plant interaction outcome and how limitation of multiple resources (e.g., light, water) modulates plant interactions at the single-species and community levels. We selected four dominant shrub species belonging to two different functional types, two Fabaceae, Cytisus galanoi and Genista versicolor, and two non-N-fixing species, Bupleurum spinosum (Apiaceae) and Horrnathophylly spinosa (Brassicaceae), and assessed shrub effects on performance of a focal understory species, Euphorbia nicaeensis, using a functional trait approach (Violle et al., 2007). We additionally estimated shrub effects on the herbaceous plant community beneath shrubs, recording the number of species and of individuals per species. Hence we tested whether effects at the species level paralleled effects at community level, and whether these effects were consistent across different shrub species and functional types. We expected that shrubs from different functional type would differ in their effects on Euphorbia performance and the understory community, and that species-specific effects on Euphorbia would be scalable to the whole community.

MATERIALS AND METHODS

Field Site and Shrub Species
Our field site was in the southern slope of the Sierra Nevada Mountains, South-East Spain. The field site is at 2040 m elevation (36°57′52″N; 03°20′12″W), slightly below the tree line. The site is dominated by shrub species like Cytisus galanoi, Genista versicolor, or Hormathophylla spinosa (Valle, 2003). The soil is an Inceptisol belonging to the Ochrept suborder (de la Rosa et al., 2001). The bedrock is mica-chissant, which determines the gentle hillsides found in the Sierra Nevada Mountains (Delgado et al., 2001). The climate is dry continental Mediterranean with a hot and dry summer (means of 17°C and 5 mm rainfall in July at Pradollano, 2500 m elevation), with 30% of the precipitation in form of snow above 1800 m. Mean annual rainfall is ~800 mm and mean annual temperature is 6°C (Delgado et al., 2001). The

1http://www.magrama.gob.es/es/red-parques-nacionales
growing season extends from June to September with peak in July.

The study was conducted on a west-oriented ~1 ha plot with 14.8% slope where we selected four shrubs species including two N-fixing species, *Cytisus galanioid* Talavera and Gibbs (Fabaceae) and *Genista versicolor* Boiss. (Fabaceae), and two non-N-fixing species, *Bupleurum spinosum* Gouan (Apiaceae) and *Hormathophylla spinosa* (L.) P. Küpfer (Brassicaceae), all common in the Sierra Nevada Mountains. Total shrub cover (a mix of species) was approximately 80% with the remaining surface covered by open areas. The cover was an order of magnitude lower for *Bupleurum spinosum*.

Effeltrich, Germany) under ambient radiation and steady-state gas exchange recorded at 3-min intervals. Instantaneous water use efficiency (WUE) was then calculated as WUE = A/E. Chlorophyll fluorescence (ΔF/Fm′) was measured using a photosynthesis yield analyzer (MINI-PAM; Heinz Walz GmbH, Effeltrich, Germany) under ambient radiation and steady-state conditions. All gas exchange and chlorophyll fluorescence traits were measured along 4 days between 8 am and 12 am under cloudless conditions.

Plant size was calculated as $S = H^2V$ where $H$ is plant height and $V$ is the number of vegetative stems. We tested the relationship between $S$ and biomass on a subset of 31 individuals collected within and outside shrubs in the same field site; both variables were positively correlated ($r = 0.95, p < 0.001, n = 31$), similar to data obtained in other species elsewhere (Prieto et al., 2011; de Bello et al., 2012). We then used plant size as a surrogate for plant biomass to avoid destructive measurements. To assess interaction intensity, the relative interaction index (RII; Armas et al., 2004) was calculated as $RII = (S_{in} - S_{op})/(S_{in} + S_{op})$; where $S_{in}$ and $S_{op}$ are plant size under shrubs and in open areas respectively.

The number of reproductive stems and the ratio of vegetative-to-reproductive stems were determined for each *Euphorbia* individual. Infrutescence volume (i.e., the volume of the reproductive plant parts containing the seeds) was calculated as a spherical ellipsoid based on the two diameters measured on the horizontal plane orthogonal to the reproductive stem. In July 2011, reproductive stems of *Euphorbia* individuals were
TABLE 1 | Functional, growth and reproductive plant traits of *Euphorbia nicaeensis* individuals and abiotic variables under four shrub species (*Genista versicolor*, *Cytisus galianoi*, *Hormathophylla spinosa* and *Bupleurum spinosum*) and in open areas.

| Functional traits | Cytisus | Genista | Bupleurum | Hormathophylla | Open area |
|-------------------|---------|---------|-----------|----------------|-----------|
| Photosynthesis rate ($\mu$mol $m^{-2}$ s$^{-1}$) | 3.77 ± 0.38 | 4.26 ± 0.66 | 4.02 ± 0.55 | 3.51 ± 0.66 | 4.01 ± 0.63 |
| Transpiration rate (mmol $m^{-2}$ s$^{-1}$) | 0.25 ± 0.05 | 0.21 ± 0.04 | 0.24 ± 0.03 | 0.21 ± 0.04 | 0.24 ± 0.04 |
| Stomatal conductance (mmol $m^{-2}$ s$^{-1}$) | 11.36 ± 2.13 | 9.53 ± 1.75 | 11.22 ± 1.21 | 9.52 ± 2.00 | 10.83 ± 1.74 |
| Instantaneous water use efficiency (E) (mmol $m^{-2}$ s$^{-1}$) | 18.94 ± 2.42 | 21.97 ± 1.75 | 18.50 ± 2.84 | 21.31 ± 2.7 | 18.75 ± 2.24 |
| C/N | | | | | |
| Growth and reproductive traits | | | | | |
| Plant height (cm) | 25.58 ± 0.99a | 23.27 ± 1.43ab | 21.94 ± 1.60ab | 20.18 ± 1.56b | 18.45 ± 1.61b |
| Vegetative stems (n.s.) | 3.42 ± 0.40 | 4.00 ± 0.43 | 2.71 ± 0.45 | 3.64 ± 0.54 | 4.70 ± 0.99 |
| Reproductive stems** | 3.08 ± 1.43a | 1.64 ± 0.51a | 2.09 ± 0.91a | 2.45 ± 1.01a | 12.20 ± 3.48b |
| Infrutescence volume** | 2.26 ± 0.67ab | 1.21 ± 0.53a | 1.37 ± 0.49a | 0.49 ± 0.17a | 4.14 ± 1.06b |
| Seed mass (n.s.) | 0.73 ± 0.23 | 0.70 ± 0.31 | 0.91 ± 0.27 | 0.80 ± 0.29 | 1.26 ± 0.23 |
| Total seeds* | 60.59 ± 6.43a | 52.09 ± 37.66a | 62.21 ± 37.94a | 44.45 ± 24.62a | 182.1 ± 9.87b |
| Abiotic variables | | | | | |
| PAR*** | 361 ± 39.53a | 344 ± 37.88a | 240 ± 54.35a | 267 ± 46.12a | 1328 ± 62.59b |
| SWC*** | 2.72 ± 0.25ab | 4.03 ± 0.78a | 1.62 ± 0.14bc | 2.35 ± 0.23bc | 0.87 ± 0.11c |
| RH*** | 60.25 ± 2.24a | 63.42 ± 2.24a | NA | 66.10 ± 2.22a | 50.65 ± 2.02b |
| T*** | 16.05 ± 0.49a | 17.04 ± 0.52a | NA | 16.26 ± 0.45a | 19.03 ± 0.57b |

Values are mean ± SE. Results of overall test are shown in parentheses: n.s., non significant, *P < 0.05, **P < 0.01, ***P < 0.001. Different letters indicate significant differences between shrubs or with open areas (Tukey HSD tests; *P < 0.05*).

**Community-Level Data**

We recorded the number of understory species and individuals by sampling 30 quadrats under each shrub species and 120 in open areas (Supplementary Table S1). For the largest shrubs, *Bupleurum spinosum* (canopy area of 4992 ± 515 cm²; mean ± SE), *Cytisus galianoi* (8898 ± 947 cm²) and *Genista versicolor* (7290 ± 1298 cm²) we sampled one 25 cm × 25 cm quadrat randomly placed within the canopy, and for the smallest shrub, *Hormathophylla spinosa* (2452 ± 263 cm²), we sampled the whole area beneath the canopy. A quadrat of the same size (25 cm × 25 cm, or size of the whole canopy in case of *H. spinosa*) was randomly sampled in a nearby open area (at least 1 m away from the shrub). Measurements were done at the peak growing season (mid July 2011).

**Statistical Analyses**

We used redundancy analysis (RDA; Canoco 4.5 package; ter Braak and Šmilauer, 2002) to assess correlation between *Euphorbia* traits in different microhabitats. RDA provides a good estimate of main trends in the data and can be considered an extension of multivariate regression for a multivariate response variable (Lepš and Šmilauer, 2003). The parametric test is replaced by Monte Carlo permutations to overcome problems with distributional characteristics. We used the forward selection procedure to identify the main environmental characteristics explaining a significant part of the variation in *Euphorbia* traits. First, we assessed the global significance of the model. Second, we used partial analyses (i.e., we included the tested factor as the only explanatory variable and the other factors as covariates) to obtain the variability explained by each individual factor and their corresponding significance (*P*-value). Tests

bagged with nylon mesh to prevent seed loss; in September 2011, once all seeds were estimated to be mature, seeds contained in the mesh bags were collected, counted and dried for 72 h at 60°C. The number of reproductive stems, number of seeds and mean seed mass were used as surrogates of plant fitness.
FIGURE 1 | Redundancy analysis, RDA, triplot ordination with recorded Euphorbia traits and model selected environmental conditions (PAR and Microhabitat type: open area and N-fixing). Response variables (in italics) are indicated by arrows and labeled as follows: \( N \), leaf nitrogen content; \( C \), leaf carbon content; SLA, specific leaf area; WUE, water use efficiency; \( A \), photosynthetic rate; RII, relative interaction index; RWC, relative water content; LDMC, leaf dry matter content; \( C/N \), leaf C to N ratio. The environmental variables are indicated by a bold red arrow for continuous variables (PAR) and triangles in the case of categorical variables (Open area, N-fixing and Non N-fixing species). Centroids of species identity (circles) and soil water (dashed line) are added as supplementary variables without any effect to the analyses. The first canonical axis explains 11.1% of variability \( (F = 6.39, P = 0.002) \) and all three canonical axes explain 18.4% of variability \( (F = 3.83, P = 0.002) \). The explanatory variables Open area, N-fixing and PAR were selected by forward selection procedure (all at \( P < 0.05 \)) but the variable Non N-fixing is also shown to help interpretation.

FIGURE 2 | Soil water content (SWC) and mean daily photosynthetically active radiation (PAR) measured under four shrub species and in open areas in the southern slope of the Sierra Nevada Mountains. Mean values \( \pm SE \) are shown. Asterisks indicate differences with open areas (Dunnett test, \( P < 0.05 \)). Horizontal lines represent mean values for open areas. Cy, Cytisus galianoi; Ge, Genista versicolor; Bu, Bupleurum spinosum; Ho, Hormathophylla spinosa. Shaded area includes non-N-fixing species.

were based on 499 random permutations. The ordination diagram as an output of the RDA shows linear correlations between quantitative and qualitative environmental variables (arrows and points, respectively) and measured Euphorbia traits (arrows).

We tested differences in environmental conditions and in functional traits of Euphorbia growing under the different shrub species and in open areas using one-way ANOVA and subsequent post hoc Tukey HSD and/or Dunnett tests. We performed a second comparison to test for differences between functional types by grouping plants under N-fixing and non-N-fixing species. ANOVA assumptions of equal variance were checked using a Levene's test and normality of the residuals using a Shapiro–Wilk test (Shapiro and Wilk, 1965); when data were not normal, non-parametric Kruskall–Wallis analyses were used followed by single pair comparisons. To avoid type I error,
pair-wise comparisons between treatments for Kruskal–Wallis tests were followed by Bonferroni corrections adjusted according to Legendre and Legendre (1998). When analyzing RII, we performed Student t-tests against a constant value (zero). Positive RII values significantly different from zero indicate net positive effects (i.e., facilitation) and negative, significantly different from zero indicate net negative effects (i.e., competition). When not significant, net effects were neutral (i.e., partial positive and negative effects have the same magnitude). Long-term soil moisture below shrubs and in open areas was compared using a general linear mixed model (GLMM) with species (including open areas) and time as fixed factors and individual shrubs (or individual open areas) as the random factor. GLM, ANOVA, Kruskal–Wallis and GLMM analysis were performed using R version 2.15.3 (R Development Core Team, 2013).

We tested for differences between microhabitats (species and functional types) regarding number of individuals per square meter in the understory using Wilcoxon matched pairs tests. Each shrub species was compared with their paired open area due to different plot sizes for big and small shrubs.

RESULTS

Redundancy analysis showed significant conditional effects of three environmental characteristics on Euphorbia growth, one continuous variable (PAR), and two categorical variables related to species identity, open areas and N-fixing ability (Figure 1). These environmental traits altogether explained 18.4% of variance. Partial effects of all selected explanatory variables were highly significant (Open areas, \( P = 0.002 \); Functional type, \( P = 0.004 \); PAR, \( P = 0.024 \)). These variables explained 5.5, 5.0, and 3.2% of variance in our dataset, respectively. The first ordination axis was positively correlated with PAR and tell apart Euphorbia individuals according to their position under shrubs (negative values) versus open areas (positive values). All reproductive traits (number of reproductive stems, infrutescence volume, total seeds and seed mass) were positively correlated to this first axis; similarly, \( \Delta F/Fm' \), LDMC and C/N were also positively correlated to PAR and open areas (Axis 1). On the other hand, nitrogen content (N), SLA and plant height were negatively correlated to the first axis of variation and were thus negatively correlated to PAR. Individually, LDMC was negatively correlated to leaf N, whereas SLA was positively correlated to leaf N. The second ordination axis shows differences between N-fixing and non-N-fixing shrub species. RWC and RII were positively correlated to this axis of variation, both being greater in Euphorbia individuals growing under N-fixing shrubs. Similarly, plant height, the number of vegetative and reproductive stems, and leaf N tended to be positively correlated to N-fixing shrubs. Finally, there was no correlation between A or WUE to any of these two axes.

Abiotic Conditions

Soil water content was higher and PAR lower under shrubs than in open areas (Table 1, Figure 2), and both SWC and PAR were greater under N-fixing than under non-N-fixing shrubs (Table 2). Microhabitats differed in mean RH (\( F = 17.51, P < 0.001 \)) and air temperature (\( F = 5.60, P < 0.023 \)). Mean air temperature was higher and RH was lower in open areas than under shrubs (Table 1; Supplementary Figure S1). Long-term soil moisture was lower in open areas than under Hormathophylla or Genista, with Cytisus showing intermediate values (\( F = 2.73, P < 0.001 \)).

Plant Functional Traits

Euphorbia leaf traits (i.e., LDMC and SLA) did not differ among shrubs, between shrubs and open areas (Table 1), or between functional types (Table 2). Organic C and total N contents were higher in Euphorbia individuals growing under shrubs than in gaps (Table 1; Figure 3). Leaf N was also higher in Euphorbia individuals growing under N-fixing than under non-N-fixing shrubs (Table 2; Figure 3). Neither shrub species nor functional type had an effect on the physiological status of Euphorbia; A, E, gs, WUE, \( \Delta F/Fm' \) and RWC did not differ among microhabitats (species or open areas, Table 1), nor between N-fixing and non-N-fixing shrubs (Table 2).

Euphorbia individuals were taller under Cytisus than in open areas (Figure 3; Table 1), and also under N-fixing than under non-N-fixing species (Figure 1; Table 2). The relative interaction index (RII) was greater under N-fixing than under non-N-fixing shrubs (\( F = 14.58, P < 0.001 \)), and did not differ from zero in N-fixing species (Genista and Cytisus) indicating neutral effects, while it was significantly negative in non-N-fixing species

| Table 2 | ANOVA (regular font) and Kruskal–Wallis (italics) results for plant growth, ecophysiological traits, and continuous environmental variables of Euphorbia nicaeensis individuals growing in contrasting microhabitats under the N-fixing shrub species (Genista versicolor, Cytisus galianoi) versus non-N-fixing species (Hormathophylla spinosa, Bupleurum spinosum) in the southern slope of the Sierra Nevada Mountains (Granada, Spain). |
|--------|-----------------|
| A      | 0.18 0.675      |
| E      | 0.02 0.878      |
| gs     | 0.00 0.966      |
| WUE    | 0.04 0.844      |
| \( \Delta F/Fm' \) | 0.24 0.625 |
| SLA    | 0.58 0.447      |
| LDMC   | 1.75 0.192      |
| RWC    | 2.61 0.114      |
| C      | 1.78 0.189      |
| N      | 13.30 0.001     |
| C/N    | 12.84 0.001     |
| Plant height | 3.95 0.050 |
| Vegetative stems | 1.44 0.230 |
| Reproductive stems | 0.00 0.972 |
| Infrutescence volume | 1.32 0.250 |
| Seed mass | 1.07 0.350 |
| Total seeds | 0.11 0.739 |
| PAR    | 5.72 0.021      |
| SWC    | 6.70 0.001      |

Variables are named according to acronyms in Table 1. Bold indicates significant differences between N-fixing and non-N-fixing shrubs (\( P < 0.05 \)).
FIGURE 3 | Continued

Plant height, leaf N content, and relative interaction index (RII) in *Euphorbia nicaeensis* plants growing under four shrub species and in open areas in the southern slope of the Sierra Nevada Mountains. Mean values ± SE are shown. Different letters indicate significant differences between shrubs (Tukey HSD tests; *P* < 0.05), and asterisks indicate differences with open areas (Dunnett test, *P* < 0.05). Horizontal lines represent mean values for open areas; note that RII = 0 in open areas by definition, in this case asterisks indicate RII is significantly different from zero (t-test, *P* < 0.01). Cy, *Cytisus galianoi*; Ge, *Genista versicolor*; Bu, *Bupleurum spinosum*; Ho, *Hormathophylla spinosa*. Shaded area includes non-N-fixing species.

Contrasting allocation patterns were observed in *Euphorbia* individuals growing in open areas versus growing under shrubs, the former having lower vegetative-to-reproductive stem ratios (*F* = 2.54, *P* = 0.05). There were no differences in these ratios between shrub species or functional types (Figure 4). Furthermore, *Euphorbia* individuals growing in open areas produced more reproductive stems and had a higher number of seeds (Table 1; Figure 3) than individuals growing under shrubs regardless of shrub species or functional type. Mean seed mass, however, was similar in individuals growing under shrubs and in open areas (Table 1).

At the community level, the four shrub species had contrasting effects on the number of individuals growing in the understory. While *Cytisus galianoi* acted as a facilitator (*Z* = 2.13, *P* = 0.033), the effect of *Genista versicolor* was neutral (*Z* = 0.19, *P* = 0.85), and *Bupleurum spinosum* and *Hormathophylla spinosa* acted as competitors—although with different intensities (*Z* = 1.91, *P* = 0.05, and *Z* = 1.96, *P* = 0.049, respectively) (Figure 5). When analyzed by functional type, N fixing species exerted neutral effects (*Z* = 1.25, *P* = 0.21) whereas non-N-fixing species exerted net competitive effects (*Z* = 2.65, *P* = 0.008) (Figure 5).

**DISCUSSION**

Our data show specific effects of each shrub species on its understory community, although functional type (N-fixing vs. non-N-fixing shrubs) was a reasonably good predictor of interaction outcome. The effects of each species were consistent at the species and community levels, and were a consequence of their functional type; although we did not measure individual shrub traits, the two non-legume shrubs exerted a net competitive effect on *Euphorbia* (species level) and on the number of individuals growing beneath them (community level) whereas the two legume shrubs presented neutral effects regardless of level. Finally, *Euphorbia* displayed a high phenotypic plasticity in response to the shrub presence.

We did not find net facilitation effects but rather evidence of strong competition (under non-N-fixing shrubs) or a partial release of competition under N-fixing shrubs (Holmgren et al., 1997; Maestre et al., 2003). At the species level, this was most evident regarding biomass, as RII was neutral under N-fixing...
shrubs (Cytisus and Genista) and negative under non-fixing shrubs (Bupleurum and Hormathophylla). At community level we recorded net competition (lower number of subordinate individuals) under Bupleurum and Hormathophylla, a neutral effect under Genista, and a facilitative effect under Cytisus. We should note that facilitative and competitive effects are species-specific and depend strongly on the species under study (Pugnaire et al., 2004; Liancourt et al., 2005; Callaway, 2007; Padilla et al., 2009); however, the fact that these effects were similar at the species (e.g., for Euphorbia individuals) and community levels (number of individuals per m²) suggest that these results are scalable, as suggested already by Schöb et al. (2013b). This parallelism in the effects at the species and community levels suggests that using a single focal species provides reliable evidence on the processes shaping the plant community.

The contrasting allocation patterns recorded in Euphorbia, along with strong correlations between PAR, SWC and reproductive traits, suggest different strategies of individuals growing under shrubs and in open areas. Plants under shrubs invested relatively more in growth at the expenses of reproduction, as reported for E. terracina by Riordan et al. (2008). These differences in biomass allocation point also to the species considerable phenotypic plasticity and its ability to cope with stress meant by high light and low water availability. Since Euphorbia seeds can germinate under both light and dark conditions (Narbona et al., 2006), seeds would germinate easily under shrubs and individuals would grow better thanks to the high water and nutrient availability. Individuals in open areas, however, must cope with drier soils, higher temperatures and lower soil nutrient content, investing relatively more into reproductive organs likely to compensate for higher juvenile or adult mortality (Al Samman et al., 2001). Although reproductive allocation has a genetic basis, it can greatly vary within species depending on environmental conditions (Karlsson and Méndez, 2005; Castellanos et al., 2014). Therefore, differences in allocation patterns are Euphorbia nicaeensis’ phenotypic responses to environmental variability under shrubs and in open areas. As the number of vegetative stems, a good index of plant age in Euphorbia nicaeensis (Narbona, 2002), was similar among microhabitats, we assume a similar age for all selected individuals ruling out age driven differences on the reproductive outputs observed.

The prevalence of competitive effects has been also reported for other Euphorbia species under dry conditions (e.g., Rinella and Sheley, 2005). Improvement of soil water availability and soil nutrients under perennial species is a major source of facilitation in semi-arid systems (Pugnaire et al., 2011). Yet, given the
interaction outcome in *Euphorbia*, regarding plant size (biomass) and plant fitness (reproductive traits), and the preference of the species for sunny places (Benedi et al., 1997), it appears that light is the limiting factor for this species (cf. Grime, 1973). Our data support this idea since reproductive output (and reproductive-to-vegetative biomass ratio) was greater in open areas than under shrubs, despite lower soil water and nutrient content. *Euphorbia* individuals grew taller under shrubs than in open areas, which are a common response to light limitation (Macek and Lepš, 2003, 2008). Nevertheless, as height of *Euphorbia* individuals also differed between shrub microhabitats and did not correlate with PAR under shrubs ($P = 0.90$), it suggests there was no etiolation but rather an increase in plant size.

Although light is an important factor driving plant interactions (Holmgren et al., 2012), facilitation through increased nutrient availability is also common in semi-arid environments, especially by N under legume species (Pugnaire et al., 1996; Reynolds et al., 1999; Gómez-Aparicio et al., 2004; Armas et al., 2008). In our case, despite strong competition for light, *Euphorbia* benefited from improved N content and higher soil moisture under leguminous shrubs, as changes in leaf N positively correlated to plant height and size (i.e., taller plants and greater RII under N-fixing shrubs). Hence, the lower competition intensity observed under N-fixing shrubs was most probably a consequence of increased resource availability. This benefit may be, however, difficult to separate from improved soil water conditions, as plant N and water uptake are usually tightly linked (Hu and Schmidhalter, 2005; Macek et al., 2009).

Recent climate models predict an increase in aridity and temperature in Southern Europe (IPCC, 2007), and particularly under Mediterranean climate, where drier conditions would be found at higher elevations as compared to current conditions (Ackerly et al., 2010). Hence, plant association patterns could be altered by increased competition between shrubs and beneficiary species (Schöb et al., 2014a). An understanding of the mechanisms that drive these interactions could help forecast future changes in structure, function and assembly of plant communities under changing climate (Callaway, 2007). We hypothesize that a competitive displacement of some shrub species and a change in shrub community structure with increasing environmental harshness may alter plant interaction patterns in dry mountain systems, leading to a community structure where the ratio between N-fixing and non-N-fixing shrubs will play an important role in the Sierra Nevada Mountains.

**AUTHOR CONTRIBUTIONS**

PM, IP, and JM conceived the study. PM, IP, JM, and NP performed field data collections. PM and IP analyzed the data and PM, IP, and FP wrote the manuscript.

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**SUPPLEMENTARY MATERIAL**

The Supplementary Material for this article can be found online at: http://journal.frontiersin.org/article/10.3389/fpls.2016.00662

**FIGURE S1** | Air relative humidity in open areas ($n = 3$) and below shrubs ($n = 3$ for each shrub species) recorded during the season 2011. Mean weekly data ± 95% cv are shown. Different letters indicate significant differences between microsites (open areas and shrub species; Tukey HSD tests; $P < 0.05$).

**TABLE S1** | List of understory species found below shrubs and in open areas and the number of individuals of each species in each microsite (under shrubs and open areas). The four shrub species are depicted in bold, and total numbers of subordinate species growing beneath them are in brackets. Table modified from Pistón et al. (2018).

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The reviewer JN declared a past co-authorship with one of the authors IP to the handling Editor, who ensured that the process met the standards of a fair and objective review.

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