Measures to Control Invasive Crayfish Species in Switzerland: A Success Story?

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Invasive crayfish species were first documented in Switzerland in the 1970s. Today, North American crayfish species dominate in most major lakes and streams in Switzerland. In combination with the crayfish plague, they pose a substantial threat to our native crayfish. Over the past 20 years, various techniques have been applied to reduce negative impacts of these invasive crayfish in Switzerland: eradication (temporary drainage or destruction of a water system, biocides), suppression (intensive trapping, electricity introduction of predatory fish) and containment (construction of crayfish barriers). Temporary drainage or filling-in of isolated ponds, in combination with calcium hydroxide application has been successful in eradicating populations of invasive crayfish. However, trapping and introduction of predatory fish led to a reduction in population density but neither method has ever caused the extinction of a population. Invasive crayfish have not yet reached crayfish barriers, therefore, long-term functionality of these barriers still needs to be proven. Nevertheless, functional controls with native crayfish have shown that barriers prevent their upstream movement. Implementation of crayfish barriers is the most promising method to protect native crayfish from displacement by invasive crayfish species. Many measures are expensive, time consuming, and show little or no success in controlling invasive crayfish. Therefore, we recommend to focus on implementing drastic measures, such as filling-in or draining of isolated waters or a combination of various methods to maximise the reduction of population size.

**Keywords:** biocides, trapping, barriers, infilling, electric-fishing, drainage, function control, migration

**INTRODUCTION**

Preventing the widespread disappearance of indigenous crayfish species (ICS) in Europe is an on-going challenge. In addition to habitat destruction by river engineering and water pollution, invasive non-indigenous crayfish species (NICS) from North America are becoming increasingly widespread in Europe and pose a major threat to ICS (Kouba et al., 2014). Their dominance over native species is reflected in their rapid reproduction (Burić et al., 2011), physical superiority and aggressiveness (Gherardi, 2006) as well as their tolerance to poor water quality (Nyström, 2002). Displacement of ICS with NICS is predicted whenever they occupy the same habitat niche (Westman et al., 2002); this is if the crayfish plague (Aphanomyces astaci) has not already initiated extinction of the native species. This fungus like pathogen belonging to the Oomycetes (Saprolegniales) was responsible for the first mass mortalities of native crayfish in Europe in the...
late nineteenth century and it continues to cause huge problems in waterways today (Holdich et al., 2009).

Various methods have already been applied and tested across the world to reduce or eradicate the negative impacts of the unwanted intruders as well as to prevent them from spreading further. These methods include: intensive trapping (Bills and Marking, 1988; Hein et al., 2007), male sterilisation (Piazza et al., 2015), the use of biocides (Cecchinelli et al., 2012), habitat destruction, the release of predators (Musseau et al., 2015), the construction of barriers (Cowart et al., 2018) or a multi-method approach combining different combinations of these methods (Hein et al., 2006; Freeman et al., 2010; Stebbing et al., 2014). The success of these measures varies and so far only attempts which fill-in isolated still waters or the use poison (Ballantyne et al., 2019) have been successful in completely eradicating populations of NICS.

In Switzerland, the three native species Astacus astacus (Linnaeus, 1758), Austropotamobius pallipes (Lereboullet, 1858) and Austropotamobius torrentium (von Paula Schrank, 1803) are threatened by the presence of three North American crayfish species. Faxonius limosus (Rafinesque, 1817) was detected for the first time in Switzerland in 1976. It was followed by Pacifastacus leniusculus (Dana, 1852) at the end of the 1980s and Procambarus clarkii (Girard, 1852) in the 1990s. F. limosus and P. leniusculus have spread rapidly across Switzerland and are, through transmission of the crayfish plague, the main reason for population extinction of ICS (Stucki and Zaugg, 2011).

The three native species in Switzerland are defined as internationally protected in Appendix III of the Convention on the Conservation of European Wildlife and Natural Habitats which was decided upon in the Bern Convention 1979. In order to protect these species, Switzerland is tackling not only reintroduction, conservation and habitat management but also containment, suppression and eradication of NICS (Hefti and Stucki, 2006).

This paper summarises and assesses the success of various measures taken so far to combat and prevent the spread of NICS in Switzerland. The results serve as a basis for the future strategy to manage invasive crayfish species in Switzerland and help other countries to implement similar measures.

**MATERIALS AND METHODS**

Cantonal fisheries departments in Switzerland provided information regarding which measures they had already taken and provided existing data about implementation and success of the methods. The data received was examined to assess whether it could be used to indicate the success of a control measure. When catch per unit effort (CPUE) values were recorded over several years or when the functionality of a crayfish barrier was tested in the field (function control) with ICS, data was assessed as “valid.” Personal assessments, measures which have started within the last 3 years and catch figures without indication of the sampling effort were rated as “insufficient” to produce valid data and are classified and referred to as “unclear” in this paper. “Success” was defined by the data which showed a reduction in CPUE by at least 75%; when no invasive crayfish were detected above a crayfish barrier as well as when there was no further evidence of a crayfish appearance after an eradication campaign over a 5 year period. If CPUE decreased only by ≤ 25% after the eradication measurements were enforced this was judged as “no success.”

**RESULTS**

Overall, nine cantons provided data regarding 40 control measures carried out at a total of 27 sites, 9 lotic and 18 lentic waterbodies (Table 1). Data on catch effort in large watersystems, such as Lake Zurich, Geneva and Neuchâtel as well as in the river Rhine, were insufficient and not included. When excluding sites which were classified as “unclear”, ~83% of eradication, 20% of suppression, and 100% of containment methods were successful (Figure 1).

**Eradication of NICS**

Temporary drainage or destruction of habitat and the use of biocides are methods used to eradicate unwanted crayfish species in Switzerland. Attempts to eradicate populations of crayfish were successful at five out of seven sites including one isolated lotic waterbody. In one case success could not be evaluated because of insufficient data and the other attempt was unsuccessful.

**Temporary Drainage or Destruction of Habitat**

A 300-m section of the small brook, Stadtbach, Baden, which is isolated by culverts, contained a population of P. leniusculus. The stream section was completely drained in January and February 2004 and 2005. Crayfish were removed by hand following drainage but even after this crayfish could still be found. For the entire summer of 2013 this section of stream was left drained again and since then no crayfish have been detected (Stucki, 2018).

In 2013, complete drainage and hand removal of crayfish at two small quarry ponds, (Steinbruch Mellikon) led to the successful eradication of a population of P. clarkii. The project was carried out in close cooperation with the amphibian managers at this site (Stucki, 2018).

There was an attempt to eradicate P. clarkii inhabiting an artificial pond in a public park (l’étang de Vidy) by draining away all of the water and treating the soil with calcium hydroxide. After treatment, the soil at the bottom of the pond was removed. One year after the procedure, crayfish were still caught in traps in the pond. The process was repeated again, but this time refuges in the banks were also concreted over (Girardet et al., 2012). No crayfish have been detected since completing the second procedure.

Similar success was also achieved with another pond (Kunzareal, Rheinfelden) inhabited by P. clarkii. In 2008, the pond was successfully restocked with A. astacus after it had been drained from winter 2006 until summer 2007 and following subsequent calcium hydroxide treatment (Stucki, 2018).

A pond populated by P. leniusculus was filled with excavated material to restore a disused gravel pit (Kiesgrube Rohr). This measure was successful in eradicating the isolated crayfish population (Stucki, 2018).
| Site name                  | Water type | Species          | Aim            | Method          | Duration                  | Change of CPUE | Success |
|--------------------------|------------|------------------|----------------|-----------------|---------------------------|----------------|---------|
| Aubach                   | Lotic      | *P. leniusculus* | Suppression    | Trapping        | Continuous since 2009     | Increase by 100% | No**    |
| Birsig                   | Lotic      | *P. leniusculus* | Suppression    | Trapping        | Continuous since 1997     | –               | Unclear* |
| Brook Near Origlio       | Lotic      | *P. clarkii*     | Suppression    | Trapping        | Since 2013                | Increase by 30% | No**    |
| Brook Near Origlio       | Lotic      | *P. clarkii*     | Suppression    | Hand catch      | Since 2013                | Increase by 30% | No**    |
| Dättwiler Weiher         | Lentic     | *P. leniusculus* | Suppression    | Trapping        | Continuous since 1997     | Decrease by 75% | Yes**   |
| Dättwiler Weiher         | Lentic     | *P. leniusculus* | Containment    | Overflow pipe   | Since 2002                | –               | Yes**   |
| Dättwiler Weiher         | Lentic     | *P. leniusculus* | Suppression    | Predatory fish  | Campagin in the 90s       | Decrease by 75% | Yes**   |
| Depotsee Bern            | Lentic     | *O. limosus*     | Suppression    | Hand catch      | 1999–2004                | –               | Unclear* |
| Egelsee                  | Lentic     | *P. limosus*     | Containment    | Overflow pipe   | Since 2013                | –               | Unclear* |
| Eisweier                 | Lentic     | *P. leniusculus* | Suppression    | Trapping        | Since 2013                | Increase by 20% | No**    |
| Etzgerbach               | Lotic      | *P. leniusculus* | Containment    | Crayfish barrier| Since 2015                | 100%            | Yes**   |
| Fischzucht Heuwies       | Lentic     | *P. leniusculus* | Eradication    | Drainage        | 2012–2014 and 2019       | –               | Unclear* |
| Greifensee               | Lentic     | *O. limosus*     | Suppression    | Predatory fish  | Campaign 2019            | –               | Unclear* |
| Katzensee                | Lentic     | *P. clarkii*     | Suppression    | Trapping        | Since 2015                | –               | Unclear* |
| Katzensee                | Lentic     | *P. clarkii*     | Suppression    | Predatory fish  | 2019                      | –               | Unclear* |
| Kiesgrube Rohr           | Lentic     | *P. clarkii*     | Eradication    | Filled-in       | Campagin in the 90s       | –               | Yes**   |
| Kunzarena Rheinfelden    | Lentic     | *P. clarkii*     | Eradication    | Drainage        | 2006/2007                | –               | Yes**   |
| L'étang de Vidy          | Lentic     | *P. clarkii*     | Eradication    | Drainage        | 2008 et 2010             | –               | Yes**   |
| Lützel                   | Lotic      | *P. leniusculus* | Containment    | Crayfish barrier| Since 2016                | –               | Unclear* |
| Mellinger Lanklagenweiler| Lentic     | *P. clarkii*     | Suppression    | Trapping        | Since 1997                | –               | No**    |
| Mellinger Lanklagenweiler| Lentic     | *P. clarkii*     | Suppression    | Predatory fish  | Since 1997                | –               | No**    |
| Mellinger Lanklagenweiler| Lentic     | *P. clarkii*     | Eradication    | Biocide         | Campaign in 2007          | –               | No**    |
| Mellinger Lanklagenweiler| Lentic     | *P. clarkii*     | Containment    | Catch basket    | Since 2000–2016           | –               | Unclear* |
| Mellinger Lanklagenweiler| Lentic     | *P. clarkii*     | Containment    | infiltration system| since 2016              | –               | Unclear* |
| Pisciculture de Saint-Victor | Lentic  | *P. leniusculus* | Suppression    | Predatory fish  | Campaign in 2020          | –               | Unclear* |
| Pisciculture de Saint-Victor | Lentic  | *P. leniusculus* | Suppression    | Trapping        | Since 2020                | –               | Unclear* |
| Pond Near Passeiry        | Lentic     | *P. leniusculus* | Suppression    | Trapping        | Since 2020                | Decrease by 100%| Yes**   |
| Pond Near Passeiry        | Lentic     | *P. leniusculus* | Suppression    | Trapping        | Since 2020                | Decrease by 100%| Yes**   |
| Pfaffnern                 | Lotic      | *P. leniusculus* | Containment    | Crayfish barrier| Since 2017                | –               | Yes**   |
| Riehensteich             | Lentic     | *P. leniusculus* | Suppression    | Trapping        | Since 2009                | Increase by 70% | No**    |
| Roulave                   | Lotic      | *P. leniusculus* | Containment    | Crayfish barrier| Planned for 2021          | –               | Unclear* |
| Rumensee                  | Lentic     | *P. clarkii*     | Suppression    | Trapping        | 2007 and 2018             | –               | Unclear* |
| Schübelweier              | Lentic     | *P. clarkii*     | Suppression    | Trapping        | Start 2002                | Decrease by 75% | Yes**   |
| Schübelweier              | Lentic     | *P. clarkii*     | Suppression    | Predatory fish  | Start 2002                | Decrease by 75% | Yes**   |
| Stadtbach, Baden          | Lotic      | *P. leniusculus* | Suppression    | Hand catch      | Since 1997                | –               | No**    |
| Stadtbach, Baden          | Lotic      | *P. leniusculus* | Suppression    | Electricity     | Campagne in the 90s       | –               | No**    |
| Steinbruch Mellikon       | Lentic     | *P. clarkii*     | Eradication    | Drainage        | 2004 and 2005             | –               | Yes**   |
| Wenkenweier               | Lentic     | *A. leptodactylus*| Suppression    | Trapping        | Since 2009                | Increase by 30% | No**    |

Le, Lentic water; L, lotic water; Data rating: * = insufficient; ** = valid.

In Fischzucht Heuwies, an attempt to completely remove a population of *P. leniusculus* by draining the fish pond from December 2012 to April 2014 was almost successful. No crayfish was found during monitoring from 2015 to 2018. In 2019, the pond was drained again to carry out construction works to reconnect the pond to the brook. In the drained pond,
12 *P. leniusculus* were found in multiple hand catches. During construction work, the drained pond was mechanically treated and the majority of the soil at the bottom of the pond was removed (∼10,000 m$^{-3}$). Since reconstruction, no invasive crayfish have been detected with traps or using eDNA. The success of the eradication measure in *Fischzucht Heuwies* is still open as it was conducted only one year before publication of this article.

**Biocides**

Liquid manure was used in an attempt to eradicate a population of *P. clarkii* living in an artificial outflow culvert of a pond (*Mellinger Tanklagerweiher*). The culvert was blocked at one end and filled with liquid manure from cows and hens and left for 48 h before being removed. Complete eradication of crayfish in the treated area was not possible using this method (Stucki, 2018).

In the 1990s, the fishing authorities planned to treat two ponds (*Schübelweiher, Rumensee*) inhabited by invasive *P. clarkii* with fenthion, an insecticide used to kill mosquitos, fleas and ticks as well as any unwanted fish and crayfish in fish breeding stations. However, the project was stopped due to protests from nature conservation organisations including the WWF and local activists about the predicted negative impacts which fenthion would have had on other aquatic life. An alternative action plan was carried out avoiding the use of harmful chemicals; the combined use of predatory fish and trapping were sufficient to regulate the invasive crayfish population and reduce their negative impacts on ecosystem (Borner et al., 1997, 1998; Frutiger and Müller, 2002).

**Suppression of NICS**

To suppress NICS mechanical removal by traps, hand-catch, electricity and introduction of predatory fish were applied. In 2 of 12 lentic waters (∼17%) trapping of *P. leniusculus* in combination with the introduction of predatory fish led to a 75% reduction in CPUE. In six of the lentic waters, data was classified as “unclear” and in four cases there was no success despite valid data. In lotic waters, three of the control methods applied were so far unsuccessful and one success was unclear.

**Mechanical Removal**

Tapping of crayfish was carried out by cantonal fishery authorities (pond near *Passeiry* and *Pisciculture de Saint-Victor*), community service workers (*Aubach, Birsig, Eisweihir, Riehenteich*, and *Wenkenweiher*), water tenants (*Dättwiler Weiher, Katzensee, Mellinger Tanklagerweiher, Rumensee, and Schübelweiher*) and by a private environmental company (brook near *Origlio*). A corresponding permit from the responsible department for fisheries is needed to be allowed to catch invasive crayfish in Switzerland. Water tenants were obliged to remove crayfish of all age groups and of both sexes and were forbidden to put any back after they had been caught. Crayfish from *Mellinger Tanklagerweiher, Schübelweiher*, und *Katzensee* were sold to restaurants to be eaten.

At sites where trapping was combined with introduction of predatory fish a reduction in CPUE of 75% was observed in *Dättwiler Weiher, Schübelweiher* and 100% in *Pisciculture de Saint-Victor*. In *Mellinger Tanklagerweiher*, no change in CPUE was observed over 19 years despite the introduction of pike. In the pond near *Passeiry*, the CPUE decreased from 10.9 to 0 from mid-July to beginning of September. At this site, artificial refuge traps (ARTs) were used in addition to “pirate” traps (Bock-Ås Ltd., Finland). However, only 0.4% of the 2,480 individuals
caught at Passeiry were trapped in ARTs; these were mainly small individuals (<8 cm total length, sex ratio 5:4).

**Electricity**

Three methods of electricity exposure were tested to try to eradicate *P. leniusculus* in a 300 m long stretch of a culvert-isolated stream (Stadtbach, Baden); direct current (electric fishing gear, 500 V, 5 A, 30 s exposure), impulse current (electric fishing gear, 1,000 V, 1 A, 100 Hz, 5 s exposure) and alternating current (fish killing gear, 40 V, 0.35 A, 50 Hz, 60 s exposure). Crayfish exhibited strong behavioural reactions to the impulse current when alternating and direct current were applied but there were no mortalities of crayfish (Stucki, 2018).

However, electrofishing devices were successful in catching *P. clarkii* in a small brook near Origlio; on average four electrofishing-campaigns were implemented each year. At this site, electrofishing, trapping and hand catch have been used in combination for 5 years. In this brook more crayfish could be caught using electrofishing compared to trapping and hand-catch during night-time inspections. Nevertheless, there has not yet been significant decrease in the number of crayfish caught so far.

**Introduction of Predatory Fish**

Pike, *Esox lucius* (Linnaeus, 1758), perch, *Perca fluviatilis* (Linnaeus, 1758), zander, *Sander lucioperca* (Linnaeus, 1758), and *A. anguilla* are all predatory species which were introduced individually or combination into seven lentic waters to reduce the occurrence of three invasive crayfish; *P. leniusculus* (Dättilwiler Wehier, Pisciculture de Saint-Victor, and pond near Passeiry), *P. clarkii* (Mellinger Tanklagerweiher, Katzensee, Schübelweiher), and *L. fluviatilis* (Greifensee).

The introduction of eel into *Pisciculture de Saint-Victor* and the pond near Passeiry was conducted in summer 2019 and results are from autumn of the same year. In two of the lentic waters, the release of predatory fish in combination with the use of trapping, led to a reduction in CPUE of 75%; after three (Dättilwilerwehier) and six years (Schübelweiher). In four cases, success cannot be assessed due to the insufficient data gathered so far (Greifensee, Katzensee, Pisciculture de Saint-Victor, and pond near Passeiry). A reduced tendency of *P. clarkii* to spread over land was reported by the local fishery department after implementation of control measures in Schübelweiher, Katzensee, and Mellinger Tanklagerweiher.

**Containment of NICS**

To prevent further spread of NICS several barriers were constructed in Switzerland. Three out of eight barriers have so far proven their success. For the other five, an assessment of success is not yet possible and have so far been classified as “unclear”.

**Crayfish Barriers in Lotic Waters**

In the Etzgerbach, a crayfish barrier was built to prevent *P. leniusculus* from migrating into the headwaters where native *A. torrentium* are found. The barrier consisted of a 30 cm free fall with a steel overhanging lip. A function control with native *A. astacus* could confirm that the barrier functioned in preventing crayfish movement upstream while also allowing migration of trout (*Salmo trutta fario* Linnaeus 1758) with a body length ranging from 11.5 to 49 cm.

Another barrier with the same design as in Etzgerbach was built in the river Lützel in order to protect the habitat of *A. pallipes* from *P. leniusculus* invasion. There has been no detection of *P. leniusculus* in traps upstream of the barrier, even 4 years after its construction.

A 40 cm high crayfish barrier made out of stainless steel with a fish passable mid-section was built with connection to the river bed in the Pfaffnern. The stainless steel created smooth surfaces that in combination with high water velocities, ≥0.65 m/s, aimed to stop *P. leniusculus* from spreading further upstream (Frings et al., 2013). PIT-tagged *A. astacus* were released downstream of the barrier to verify if crayfish could overcome the barrier. No released *A. astacus* was detected upstream of the barrier during the 8 month study.

**DISCUSSION**

**Eradication of NICS**

**Temporary Drainage or Destruction of a Habitat**

Complete destruction of an isolated waterbody has led to a successful eradication of the invasive crayfish population in Kiegrube Rohr and is the most promising method for future eradication attempts. Although drainage of waterbodies has been successful (Steinbruch Mellikon and l'étange de Vidy), if the water is only drained for a few months, there is always the risk that some crayfish will survive in burrows or humid places and can rebuild.
a stock (Fischzucht Heuwies). Treatment with calcium hydroxide can in this case increase success possibility (l'étange de Vidy and Künzereal Rheinfelden).

It is difficult to definitively state that there are no crayfish inhabiting a particular waterbody after an eradication effort. Eradication is only considered successful after 2–5 years of monitoring without a crayfish detection (Peay et al., 2006). Trapping, night inspections and eDNA testing should be used in combination to effectively evaluate the presence of crayfish.

**Biocides**

Although the use of biocides is considered the cheapest and most efficient method to eradicate unwanted crayfish populations (Manfrin et al., 2019), they have never been used in Switzerland, with the exception of liquid manure. Examples from Scotland (Peay et al., 2019) and Sweden (Ljunggren and Sundin, 2010) demonstrate that the use of biocides can be successful in eradicating NICS in isolated waters systems and that non-target invertebrates start recolonisation within a month after treatment.

**Difficulties in Eradicating NICS**

The extinction vortex and minimum viable population density is species-specific and depends on predation, reproduction strategy and environmental factors-food source, habitat size, and disease (Fagan and Holmes, 2006). In theory, only a single male and female are needed to successfully rebuild a population. One study showed that recolonisation could theoretically begin with as little as eleven individuals in an isolated water of 1 km², this figure was calculated from walking distances of male P. leniusculus during breeding season (Peay, 2001). With such a small original population there can be problems with inbreeding which can result in low genetic diversity, meaning an entire population could more easily be wiped out by disease or by a natural disaster. Establishment of all populations of F. limosus in Europe can be traced back to the introduction of only 90 individuals (Filipová et al., 2011), underlining the fact that even if only a few individuals remain after a control measure this can be sufficient enough to build up the population again (Henttonen and Huner, 1997). For marbled crayfish (Procambars virginalis) Lyko, 2017), a species which reproduce by parthenogenesis, it is theoretical even possible to rebuild a population with only one individual (Ercoli, 2019). In this case, the probability of extinction is again much higher due to environmental factors, predation and failure to breed.

**Suppression of NICS**

**Mechanical Removal**

At sites where mainly adult crayfish were caught in traps, no reduction in population size was observed (Aubach, Wenkenwein Riehen, Eisweier, Rihenteich). According to our data and in agreement with previous studies, the use of various techniques to reduce the population size clearly achieves better results than a single control approach (Manfrin et al., 2019). Species with high fecundity and early maturity react undesirably to harvest control measures because the niche of the removed adult crayfish is quickly taken over by younger individuals (Zipkin et al., 2009). This fact can explain the increase in CPUE at Aubach, Eisweier, Rihenteich, Wenkenweiner, and in the brook near Origlio. Evaluating whether NICS eradication measures have been successful can take several years (Peay, 2001). Therefore, in order to prevent the population from rebuilding, the use of traps and other control methods must continue even when no more animals are caught.

In the Passerey, the CPUE decreased to zero after one year of intense trapping with “pirate” traps and ARTs as well as after the introduction of eels. The treatment was only conducted for one year so success cannot yet be guaranteed. Trapping was stopped when no more crayfish were caught with traps. Traps have been shown to only catch a minority of the population (Chadwick et al., 2020), therefore, it can be assumed that there are still some crayfish occurring in the Passerey and so trapping should be continued to avoid population numbers bouncing back. There was limited success to reduce invasive crayfish populations using ARTs in the Passerey. In a typical upland river in the south-west of England more juvenile and female P. leniusculus were caught in ARTs than in conventional traps indicating the advantage of applying this method (Green et al., 2018). The poor catch rate in ARTs in the Passerey could be because there are many naturally existing refuges already present in the banks, which makes the traps a less attractive refuge for the crayfish. Another reason for the lack of success of ARTs in Switzerland could be the fact that the number of crayfish was quickly lowered by the eels introduced at the same time as the traps.

For Birsig, Katzensee, Greifensee, and Rumensee no data is available regarding effort and catch numbers: therefore, it is not possible to evaluate success in suppressing these populations of NICS. For this reason it is important to use CPUE values or the capture-mark-recapture (CMR) technique to estimate the population size and assess the success of the method (Zimmerman and Palo, 2011).

Invasive crayfish fishing and sale has never led to eradication of a NICS population in Switzerland. The creation of a culinary market for invasive crayfish promotes illegal stocking and means that the population of invaders is maintained instead of eradicated (Nuñez et al., 2012). This issue is also highlighted in the IAA Gotland Resolution, which was formulated at the “IAA Gotland 2019 Crayfish conference” (Edsman, 2019).

**Electricity**

In Switzerland, the use of electricity in an isolated section of the Stadlbach, Baden was not effective in eradicating P. leniusculus. However, the use of repeated high intensity (69 W, direct current 1,600 V, 57.8 A, at 7 Hz) shocks resulted in high mortality (86–97%) of P. leniusculus inhabiting a stony headwater stream in England (Peay et al., 2014). In the treatment in England, they used 1.6 V and amps eleven times higher than in Stadlbach, Baden. This could be the reason for the failure to eradicate the invasive population at this site.

In a small stream (brook near Origlio), even the use of electrofishing combined with night inspections and the use of traps, could not reduce the catch number of P. clarkii over 5 years. One reason for this could be due to the specific life-history traits
of *P. clarkii* including its high fecundity and early maturity which allows fast population recovery (Chucholl, 2011).

**Introduction of Predatory Fish**

In Switzerland, introduction of predatory fish in combination with the use of traps has been successful in reducing recorded overland movements of *P. clarkii*. However, it is assumed that the population size will grow again if trapping is stopped and the number of predatory fish decrease (Paragamian, 2010). The use of native predatory fish is regarded as a good way to reduce the number of juvenile crayfish and complements trapping which removes reproducing adults (Elvira et al., 1996; Aquiloni et al., 2010; Musseau et al., 2015).

Non-native predators can also reduce crayfish densities (Miyake and Miyashita, 2011). Thus, the removal of invasive predatory fish, namely, pike, perch or catfish, from conservation ponds, can lead to an increase in the population size of invasive crayfish; consequently, causing negative effects on other animal and plant species. The release of native predatory fish in previously uninhabited waters can also have a negative impact on the other species present, including amphibians (Braitha et al., 1996; Hecnar and M’Closkey, 1997). Since invasive crayfish also have a negative influence on native plants and animals, the pros and cons of each control method must be considered to determine which will be most beneficial for each site.

Catching of predatory fish in waterways with NICS should be prohibited or regulated. If recreational fishing is allowed in certain ponds, new groups of predatory fish should be regularly introduced to the water system to ensure that their numbers are kept high; therefore, maximising the effect of predatory fish on crayfish.

**Containment of NICS**

Preventing the further spread of NICS through artificial or natural barriers is an important tool for invasive species management. The use of artificial barriers or modification to existing structures prevents the further spread of invasive crayfish and is a more cost-effective method compared to carrying out never-ending stock control when crayfish have invaded new sections of a lotic water. In Switzerland, the erection of barriers is part of the national strategy to prevent NICS from interfering with further waters (Stucki and Zaugg, 2006).

The crayfish barriers in the *Etzgerbach*, *Lützel*, and *Pfaffnern* which are mentioned in this paper were so far able to prevent upstream movement of crayfish. Barriers are the only known way to stop the natural spread of NICS within a water system. On the other hand, crayfish barriers can impede the migration of poor swimming fish; therefore, each new case a balance of interests has to be made as to whether fish migration or the prevention of the spread of NICS is more important (Krieg and Zenker, 2020). Most crayfish barriers will be constructed to protect side waters which are mainly inhabited by trout. Functional controls carried out at the barrier in the *Etzgerbach* highlighted that marked trout were able to overcome the crayfish barrier here; indicating that trout are not negatively affected by artificial crayfish barriers.

Despite the lack of function control in *Mellinger Tanklagerweiher*, the use of an infiltration system is a promising method to prevent the spread of NICS from a pond into an adjoining brook as it is not physically possible for crayfish to dig through this infiltration system.

**Combination of Different Methods**

According to the results of this study, success in suppression of an NICS population can be better achieved by using a combination of several control methods (electricity, ARTs, hand-catch, traps, and predatory fish) rather than applying just one method on its own. The reduction in size of populations of NICS was achieved exclusively in lentic waters, when combining the use of traps with the stocking of predatory fish. Predatory fish have a larger impact on preventing population growth rate than traps because they target the offspring which are responsible for future population growth. Traps on the other hand have a larger impact on breeding adults and they often miss juveniles which are small and can escape through gaps in the traps (Hein et al., 2006). The selectiveness of each method means that they should be used in combination to be most effective. If the majority of adult crayfish are caught in traps, there can be an increase in reproduction in the population whereby females produce more eggs at a younger age and there is a higher rate of survival of juveniles because of lower intraspecific predation and increased food availability (Momot, 1998).

**Data Collection**

In this study, data showed that no conclusions could be drawn about the success of the control measures in ~43% of the sites (*n* = 12). However, in 57% of treated waters (*n* = 16), the data could be used to assess the success of the method and whether its continuation is appropriate.

The meticulous collection of catch data, in particular the determination of the CPUE, is mandatory in order to assess the growth of a population over time, thus the success of a control measure (Schwarz and Seber, 1999). By catching and measuring all age groups, it is also possible to make an assessment of the number of reproducing individuals and highlight any visible population growth trends (Rabeni et al., 1997; Paillisson et al., 2011). Another technique to evaluate population size is CMR which is a promising option in smaller ponds (Coignet et al., 2012).

**Conclusion**

It is more effective to completely eradicate an invasive species with drastic measures as soon as it is found inhabiting a water rather than continuously reducing population density by suppression or isolating a population by containment (Simberloff 2014). Populations of NICS should be removed as soon as possible to avoid further damage and destruction of the habitat for native species. If the use of biocides is not possible, it is recommended to drain or destroy water bodies whenever possible to eradicate populations of NICS. Suppression and containment methods can be used to minimise the negative effects which NICS have on a habitat.
A combination of trapping and introduction of predatory fish has proven successful as a suppression multi-method approach. Crayfish barriers are the only way to contain an established NICS population to ensure that they are isolated from invading further upstream. Global warming may pose further problems as invasive species are advantaged by the increasing water temperatures, meaning they will be able to spread even faster and colonise previously inappropriate habitats (Rahel and Olden, 2008).

It is important that measures to control invasive species are carried out and agreed nationwide and across countries to prevent spread from places where no measures are taken. This is always a difficult issue to overcome when trying to control an invasive species of animal or plant, as land or municipal boundaries set by humans do not apply to them (Fernandes et al., 2019; Beaury et al., 2020). The development of a common strategy based on the experience gained to date can significantly increase success of invasive species eradication as well as saving money, the environment and the species inhabiting it.

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DATA AVAILABILITY STATEMENT

The original contributions presented in the study are included in the article/supplementary materials, further inquiries can be directed to the corresponding author/s.

AUTHOR CONTRIBUTIONS

RK, AK, and AZ conceived the ideas and designed the methodology. RK collected and analysed the data. AK and AZ led to the writing of the manuscript and contributed critically to the drafts and all authors gave final approval for publication.

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Conflict of Interest: The authors declare that the research was conducted in the absence of any commercial or financial relationships that could be construed as a potential conflict of interest.

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