The variability in the biology of living organisms is poorly appreciated in toxicology. However, accumulating evidence indicates that sex-differences modulate toxicokinetics and toxicodynamics from cellular/molecular to whole animal levels resulting in different toxic responses of living organisms to xenobiotics exposure. In order to investigate the influence of sex in inorganic mercury (Hg) exposure, male and female Wistar rats were exposed to 0.5, 1.0 and 1.5 mg Hg/kg body weight orally as HgCl₂ twice a week for 12 weeks. Higher Hg levels in the females (except heart) as compared to males were observed in the animals. At the highest dose of inorganic Hg, female renal Hg content was 3.3 times higher than that of the males. Mixed sexual dimorphism characterised circulating-lipid- and organ-lipid lipotoxic and non-lipotoxic dyslipidemia. The highest dose of inorganic Hg, induced hypercholesterolemia in the males as opposed to hypocholesterolemia in the female. Plasma and erythrocyte free fatty acids increased in both sexes, although the increase was more pronounced in the male. Reverse cholesterol transport was inhibited in the male at the highest dose of Hg, whereas female HDL became enriched with cholesterol. Female erythrocytes had all their lipids increased, whereas only male erythrocyte triglyceride increased. Brain cholesterol and phospholipids, and splenic phospholipids were depleted in both sexes. Our findings indicate that inorganic Hg exposure appears to affect Hg and lipid kinetics differently in both sexes, thus underscoring the need to develop sex-tailored approaches in the treatment of metal toxicosis and its metabolic outcomes.
forms, each with a specific toxicity profile: Elemental or metallic mercury (Hg$^0$), inorganic mercury and organic mercury [5–8]. Elemental Hg exists as a liquid at room temperature and can be released easily into the atmosphere because of its high vapour pressure. Inorganic Hg is found in 2 oxidation states: mercurous (Hg$^{2+}$) and mercuric (Hg$^{+}$). The mercuric state may also form organic compounds such as methyl-, ethyl-, phenyl- or dimethylmercury [5–8]. The toxicity profiles, biological behaviour, toxicokinetics, toxicodynamics as well as other outcomes of Hg, depend on their chemical forms. While elemental and organic Hg are lipid-soluble and their toxicity profiles are well represented in the literature, the low liposolubility, low absorption and low passage through the blood-brain-barrier (BBB) of inorganic Hg have resulted in the paucity of data about the toxic responses associated with exposure to this inorganic form of Hg [1].

About three decades ago, Nielsen [9], predicted that future human exposure to inorganic Hg would probably lead to a few individuals occupationally exposed to high levels and much larger populations exposed to low or very low levels from dental fillings or food items containing inorganic Hg. Increased use of inorganic Hg as drugs, skin lightening creams and germicides has been observed over the years thus resulting in human exposure [5–8]. Non-medical uses of inorganic Hg include wood preservative, photographic intensifier, dry battery depolariser, tanning agent for leather and catalyst in the manufacture of chemicals such as vinylchloride [10]. In recent times, one anthropogenic activity (artisanal gold mining in which Hg is used to separate gold from lead), has witnessed a surge in the entry of women and children resulting in chronic exposure of entire households to relatively high levels of inorganic Hg [11,12].

Exposure to inorganic Hg is associated with a wide range of toxic end-points including neurological [8], renal [13], hepatic [14], gastro-intestinal [15], reproductive [16], cardiovascular [17] and developmental [18]. Most of these studies used male animals in delineating these toxic end-points. In addition, biomarkers of toxicity used were skewed towards oxidative stress and anti-oxidants. However, epidemiological studies have recognised lipidomics as an independent risk factor in the pathogenesis and progression of an array of pathologies ranging from atherosclerosis, cardiovascular disease to cancer [19]. There is also increasing evidence that environmental factors/contaminants (most especially heavy metals), are at the heart of this dyslipidemia in both humans and animals [20–24]. Our study provides a model to investigate sexual dimorphism in the toxicokinetics and toxicodynamics of orally administered low dose inorganic Hg in the rat using lipidomics as metrics of toxicokinetics and outcomes. The findings of higher levels of Hg in the organs of the females as compared to males, as well as different patterns of dyslipidemia in male vs. female and vice versa, provide ample evidence that sexual dimorphism is a common signature of the toxic responses to environmentally relevant low dose inorganic Hg exposure in the rat.

2. Materials and methods

2.1. Chemicals

Mercury chloride was purchased from Sigma-Aldrich, Missouri, USA.

2.2. Animals and treatment

Experimental protocols were conducted in accord with guidelines of the Institutional Animal Care and Use Committee and were approved by the Animal Ethical Committee of the Department of Biochemistry, Federal University of Agriculture, Abeokuta, Nigeria (FUNAAB/CBS/BCH/050269).

Sixty four Wistar rats (32 males and 32 females) bred in the College of Veterinary Medicine, Federal University of Agriculture, Abeokuta, Nigeria with body weights between 150 and 250 g (8–10 weeks old), were used for the study. They were housed in an animal room with normal controlled temperature (22 ± 2 °C) and a regular 12 h light-dark cycle (06:00–18:00 h). They were allowed 14 days to acclimatise before the commencement of inorganic Hg exposure. The animals were maintained on a standard pellet diet and had access to water ad libitum.

Animals in each sex group were divided into 4 groups of 8 animals each. While 2 groups served as control and received distilled water, the remaining groups (3 groups in each sex group) were exposed to Hg (0.5, 1.0 and 1.5 mg Hg/kg body weight) as mercuric chloride twice a week orally (by gavage) for 12 weeks (Fig. 1). These Hg doses and frequency were chosen based on previous studies [25–27]. At the end of Hg exposure, blood was collected from the abdominal arteries of the animals into heparinised tubes under light ether anaesthesia after an overnight fast. Liver, kidney, heart, lung, brain and spleen were also removed from the animals for Hg and lipid analyses. An aliquot of the blood samples was taken for Hg determination while the remaining blood samples were centrifuged to separate plasma and red blood cells. All samples were stored at –20 °C until analysed.

2.3. Mercury determination

A portion of the frozen organs (~200 mg) and whole blood (0.2 ml) were digested in nitric and sulphuric acid mixture (1:3, v/v). Total Hg was determined using Inductively-Coupled Plasma Spectrometry-Mass Spectrometry (ICP-MS) (Agilent 7700, U.S.A).

One blank and one certified reference (NIST Multielement Stock Solution, Inorganic Ventures, Christianburg, VA, USA), as well as samples spiked with known concentrations of the reference material were included in each batch of digestion. Calibration curves were obtained using 6 points with the certified standard. After each analytical run (usually 20 samples), calibration curves were obtained again to check for linearity and replication. A mean recovery rate of >95% was obtained for Hg and other elements after two determinations. Results are expressed as μg Hg/ml of blood and μg Hg/g wet weight of the organs.

2.4. Biochemical analyses

2.4.1. Plasma and lipoprotein lipid profiles

Determination of the major lipids (cholesterol, triglycerides, phospholipids and free fatty acids) in plasma and lipoproteins followed established procedures. Details of these have been given in our earlier studies [20–22,28,29].

2.4.2. Organ and erythrocyte lipid profiles

Lipids were extracted from the organs (liver, kidney, heart, lung, brain and spleen) as described by Folch et al. [30] while extraction of erythrocyte lipids followed the procedure described by Rose and Oklander [31]. After washing with 0.05 M KCl solution, aliquots of the lipid extracts were then used for the determination of lipid profiles. Details of these are given as reported earlier [20–22,28,29].

2.5. Statistical analysis

Results are expressed as mean ± SEM. Two way analysis of variance (ANOVA) followed by Tukey’s test and pairwise comparisons were used to analyse the results with p < 0.05 considered significant.

3. Results

Hg concentrations in the tissues of the animals are shown in Table 1. Irrespective of the sex, no appreciable amount of Hg was found in the blood of the animals. In contrast to blood, Hg accumulation in other soft tissues exhibited different patterns characterised by sexual dimorphism and organ specificity. While there was no uptake of Hg by the brain, a slight but significant uptake of Hg was observed in the spleen of female animals exposed to 1.0 mg Hg/kg body weight. Cardiac Hg uptake was saturated in both sexes on the exposure of the animals to 0.5 mg Hg/kg
controls respectively. Pulmonary Hg uptake peaked at 4.33 μg/g tissue in male animals exposed to 0.5 mg Hg/kg body weight followed by a 94% decrease at the highest dose of inorganic Hg. The female animals, on the other hand had their pulmonary Hg increased from 0.10 ± 0.01 μg/g to 0.83 ± 0.17 μg/g tissue in a dose-dependent manner. Hepatic Hg uptake was also characterised by sexual dimorphism, with female picking up more Hg than their male counterparts. In contrast to other organs, the kidney bioaccumulated more Hg and at a faster rate in both sexes. While the renal accumulation of Hg in male animals was not dose-dependent, a dose-dependent accumulation of Hg was observed in the kidney of the female animals. At the highest dose of inorganic Hg, the female accumulated 3.3 times more Hg than their male counterparts. Pairwise comparison reveal that female organs tended to take up more Hg than the males and the uptake appears to be dose-dependent.

Plasma lipid profiles of the animals are depicted in Fig. 2. Hg-induced dyslipidemia was characterised by hypercholesterolemia in the male animals at all the doses while hypocholesterolemia (49%) was observed in the female with 1.0 mg Hg/kg body weight (Fig. 2A). Hypertriglyceridemia (Fig. 2B) in both sexes was the hallmark of inorganic Hg exposure. Comparatively, this was more prevalent at lower doses of inorganic Hg in the male than the female. In contrast to cholesterol and triglycerides, inorganic Hg exposure did not have any effect on plasma phospholipids of the female animals whereas, at the highest dose of inorganic Hg, plasma phospholipids of the male animals was twice that of their female counterparts (Fig. 2C).

The response of reverse lipid transport in the animals to the presence of inorganic Hg is depicted in Fig. 3. In male animals, there was a gradual decrease in HDL cholesterol and phospholipids, even though the decrease was not dose-dependent (Fig. 3A and C). Although HDL cholesterol and phospholipids also decreased in the female animals, the

Table 1
Mercury concentrations in the tissues of the animals.

| Compartments     | Control               | 0.5     | 1.0     | 1.5     |
|------------------|-----------------------|---------|---------|---------|
|                  | Male/Female           | Male/Female | Male/Female | Male/Female |
| Brain Hg μg/g    | 0.05 ± 0.01a/0.06 ± 0.01e | 0.06 ± 0.01a/0.06 ± 0.01e | 0.06 ± 0.07a/0.06 ± 0.01e | 0.12 ± 0.02b/0.09 ± 0.01b |
| Liver Hg μg/g    | 0.08 ± 0.01a/0.20 ± 0.03b | 0.05 ± 0.01a/0.61 ± 0.02e | 0.17 ± 0.01a/1.37 ± 0.18e | 0.20 ± 0.01b/0.89 ± 0.06d |
| Kidney Hg μg/g   | 0.14 ± 0.01a/0.52 ± 0.03c | 29.21 ± 0.88c/22.58 ± 1.75c | 17.47 ± 0.21c/43.83 ± 4.91f | 23.78 ± 1.59e/77.89 ± 7.71f |
| Heart Hg μg/g    | 0.16 ± 0.04a/0.20 ± 0.01b | 0.75 ± 0.01c/0.75 ± 0.01c | 0.75 ± 0.01c/0.75 ± 0.01c | 0.66 ± 0.03d/0.42 ± 0.03c |
| Spleen Hg μg/g   | 0.08 ± 0.01a/0.03 ± 0.01b | 0.07 ± 0.01c/0.08 ± 0.01e | 0.06 ± 0.01c/0.17 ± 0.01d | 0.09 ± 0.01c/0.1 ± 0.01e |
| Lung Hg μg/g     | 0.03 ± 0.01a/0.03 ± 0.01b | 4.33 ± 1.77e/0.1 ± 0.01b | 0.17 ± 0.01c/0.34 ± 0.04c | 0.27 ± 0.01c/0.83 ± 0.17f |
| Blood Hg μg/ml   | 0.04 ± 0.01a/0.04 ± 0.01c | 0.04 ± 0.01a/0.02 ± 0.01c | 0.02 ± 0.01c/0.04 ± 0.01c | 0.04 ± 0.01b/0.07 ± 0.02e |

Each value represents the mean ± S.E.M. of 8 rats. Values within a row with different alphabets are significantly different from each other, p < 0.05, compared to control male.

Body weight. Doubling the Hg dose did not produce any further uptake of Hg. A tripling of the inorganic Hg dose, however, resulted in a decrease in cardiac Hg content (12% in male and 44% in female). In spite of this decrease at the highest dose of inorganic Hg, cardiac Hg contents were still 4 and 2 times higher than that of male and female controls respectively. Pulmonary Hg uptake peaked at 4.33 ± 1.77 μg/g tissue in male animals exposed to 0.5 mg Hg/kg body weight followed by a 94% decrease at the highest dose of inorganic Hg. The female animals, on the other hand had their pulmonary Hg increased from 0.10 ± 0.01 μg/g to 0.83 ± 0.17 μg/g tissue in a dose-dependent manner. Hepatic Hg uptake was also characterised by sexual dimorphism, with female picking up more Hg than their male counterparts.

Fig. 1. Experimental protocol and animal treatment.
highest dose of inorganic Hg resulted in an increase in the two lipids in this lipoprotein. In both control and inorganic Hg-exposed animals, HDL phospholipids was twice as much in female animals than their male counterparts (Fig. 3C). In contrast to cholesterol and phospholipids, inorganic Hg-induced changes in HDL triglycerides exhibited a sinusoidal pattern—a decrease, followed by an increase and a further decrease (Fig. 3B).

VLDL lipid profile is depicted in Fig. 4. While exposure to the 0.5 mg Hg/kg body weight did not seem to have any effect on VLDL cholesterol in both sexes (Fig. 4A), the remaining two doses of inorganic Hg (1.0 and 1.5 mg Hg/kg body weight) seem to modulate VLDL cholesterol in both sexes differently. While the 1.0 mg Hg/kg body weight increased male VLDL cholesterol by 28%, female VLDL cholesterol was decreased by 52%. Increasing the inorganic Hg dose to 1.5 mg Hg/kg body weight
resulted in further increase in male VLDL cholesterol whereas the increase observed in female VLDL cholesterol brought their cholesterol back to control level. In contrast to VLDL cholesterol, exposure to all the inorganic Hg doses resulted in a depletion of VLDL triglyceride in both sexes (Fig. 4B). The depletion was gradual and continuous/dose-dependent in the female whereas there was a rebound to higher VLDL triglyceride in the male animals; although at the highest Hg dose, male VLDL triglyceride was just 67% of control while female VLDL triglyceride was 24% of their control female counterparts. With exposure to the 0.5 mg Hg/kg body weight, female VLDL phospholipid phospholipid attained a peak of about 250 mg phospholipid (Fig. 4C). Doubling the inorganic Hg dose resulted in a depletion of about 100 mg phospholipid. Tripling the inorganic Hg dose resulted in a further depletion of this lipoprotein’s phospholipid by another 40 mg. In contrast however, male VLDL phospholipid increased from 58.02 ± 4.56 to 85.05 ± 8.90 mg/dL on exposure to 0.5 mg Hg/kg body weight. Doubling and tripling the inorganic Hg doses did not result in any appreciable increase as the male VLDL phospholipid content peaked at about 100 mg.

Erythrocyte lipids of the animals are shown in Fig. 5. There was a significant increase in all the three lipids in the erythrocytes of female animals. Although the increase was not dose-dependent, the highest increase in each lipid was observed in female animals exposed to the highest dose of inorganic Hg. While a 2-fold increase was observed in cholesterol (Fig. 5A), a 3-fold increase was observed in phospholipids (Fig. 5C). A 41% increase was however observed in erythrocyte triglycerides (Fig. 5B). In contrast to female animals, the three doses of inorganic Hg only up-regulated erythrocyte triglycerides in male animals in a non-dose-dependent manner (Fig. 5B). Furthermore, the 1.0mgHg/kg body weight of inorganic Hg in the male animals decreased erythrocyte cholesterol by 46% while the 1.5mgHg/kg body weight dose of inorganic Hg decreased erythrocyte phospholipids by about 33%.

Erythrocyte membrane (EM) lipid profiles as a result of exposure to inorganic Hg are shown in Fig. 6. Cholesterol was depleted 77, 50 and 73% respectively in EM of male animals whereas the decrease amounted to 73% in female animals on exposure to the highest dose of inorganic Hg (Fig. 6A). As indicated in Fig. 6B, EM triglyceride was depleted in both sexes by exposure to inorganic Hg. EM phospholipid (Fig. 6C) in both sexes responded with a decrease on exposure to inorganic Hg. Comparatively, the decrease was more pronounced in the female animals than their male counterparts. As depicted in Fig. 9A, only the highest dose of inorganic Hg resulted in an appreciable increase in brain triglyceride content in both sexes. While the increase amounted to 125% in male animals, there was a 75% increase in the female. Hepatic triglyceride increased in male animals exposed to 1.0 and 1.5 mg/kg body weight of inorganic Hg whereas the lowest dose of inorganic Hg increased hepatic triglyceride in female animals (Fig. 9B). On exposure of the animals to 0.5 mg/kg body weight of inorganic Hg, renal triglyceride decreased by 50% in both sexes (Fig. 9C). Increasing the Hg dose brought back renal triglyceride to control values in the female while a 2-fold increase was observed in the male animals. With a further increase in Hg dose, renal triglyceride in both sexes was not significantly different from controls (p > 0.05). Inorganic Hg exposure did not affect cardiac triglyceride in female animals (p > 0.05) (Fig. 9D). In contrast however, the 0.5 mg/kg body weight of inorganic Hg increased cardiac triglyceride by 25% in the male animals while the 1.5 mg/kg body weight of inorganic Hg decreased cardiac triglyceride by 25%. Irrespective of the dose, pulmonary triglyceride was not affected in the male by Hg exposure whereas a 6-fold increase in pulmonary triglyceride was observed in the female animals exposed to the highest dose of inorganic Hg (Fig. 9E).

As indicated in Fig. 9F, exposure to 1.0 mg/kg body weight of inorganic Hg resulted in depletion of splenic triglyceride (54% in the male and 41% in the female). Increasing the Hg dose restored splenic triglyceride to control values in the female whereas splenic triglyceride in the male attained 67% of control value.

Depletion of brain and cardiac phospholipids was the hallmark of inorganic Hg exposure in the animals irrespective of the sex. (Fig. 10A and D). Comparatively, the depletion was more pronounced in the female animals compared to their male counterparts. In contrast to both brain and heart, phospholipidosis characterised inorganic Hg effect in both sexes in the liver (Fig. 10B) whereas pulmonary phospholipidosis

Fig. 5. Effects of inorganic mercury on erythrocyte lipid concentrations of the animals; cholesterol (A), triglyceride (B) and phospholipid (C). Each bar represents the mean ± SEM of 8 rats. Bars with different alphabets are significantly different from each other at p < 0.05 compared to their control counterpart.
was only observed in male animals (Fig. 10E). Inorganic Hg exposure did not affect renal phospholipids in both sexes (p > 0.05) (Fig. 10C). A transient, but significant phospholipidosis was observed in the spleen of female animals on exposure to the 0.5 mg/kg body weight dose of Hg (Fig. 10F). Doubling and tripling the Hg doses however resulted in depletion of splenic phospholipids in both sexes.

Cholesterol:phospholipid ratios in different compartments of the animals on exposure to inorganic Hg are indicated in Table 2. Three distinct patterns could be recognised. In the first pattern, this ratio decreased in male animals but increased in their female counterparts. This pattern was found in the cardiac and pulmonary compartments. In the second pattern, the ratio increased in males, but decreased in the females as exemplified by renal and hepatic compartments. Erythrocyte and plasma cholesterol:phospholipid ratios were epitomes of the third pattern in which the ratio both decreased and increased in the males, but decreased in the females. In the EM however, cholesterol:phospholipid ratio both decreased and increased in the male while increasing in the female.

4. Discussion

The findings of this study indicate that sexual dimorphism not only characterised the toxicokinetics of sub-chronic inorganic Hg exposure in the rat, the attendant toxicodynamics as reflected by lipotoxic and non-lipotoxic dyslipidemia, was also sexually dimorphic. While only one pattern of sexual dimorphism could be recognised in Hg toxicokinetics, sexual dimorphism in Hg toxicodynamics (dyslipidemia) exhibited different patterns.

The deposition of a larger fraction of the entire body burden of Hg in the kidney followed by a smaller fraction in the lungs, heart, and liver as observed in this study, epitomises the “normal” pattern of toxicokinetics of inorganic Hg exposure reported in the literature [7,9,10,32,33]. However, the attainment of a higher concentration of Hg in the kidney (and to a smaller extent in the lungs, liver and spleen) of female animals compared to males, was the hallmark of sexual dimorphism in Hg toxicokinetics observed in this study. This observation was first reported by Friberg [34] following chronic subcutaneous administration of HgCl₂ to rats. Female rats not only accumulated more Hg in their kidney than their male counterparts, the rate of renal clearance was also slower in females than males [34]. This observation of Friberg lost traction as a result of Minamata epidemic (Hg poisoning as a result of exposure to methyl Hg) which broke out the same year in Japan. Since then, scientific efforts were concentrated on this organic form of Hg culminating in a lacuna in information regarding inorganic Hg. However, in related studies a few years later using methyl Hg, similar observations were also made [33,35–37]. In another study using inorganic Hg and F344 rats, a similar magnitude of sexual dimorphism in Hg toxicokinetics was observed [25]. These then indicate that irrespective of the Hg species administered, sexual dimorphism seems to be a common signature of Hg toxicokinetics. At lower doses of methyl Hg in these studies, the female-to-male ratios of Hg concentrations were less than 1 whereas at higher doses, the ratio exceeded 1. Data in Table 1 in this study indicate
that the ranges of female-to-male ratios of this Hg accumulation varied from organ to organ. While this ratio was less than 1 for the kidney and lungs on exposure of the animals to 0.5 mg Hg/kg body weight, it ranged from 2.5 to 3.3 (kidney), 4.5 to 12.2 (liver), 2.4 to 3.0 (lungs) and 1.1 to 2.8 (spleen) when the rats were exposed to 1.0 and 1.5 mg Hg/kg body weight respectively. Although these observations are consistent with what obtains for methyl Hg and are indicative of slower rate of clearance of Hg from these organs, it remains to be elucidated whether the uptake mechanism was saturable or not. Results of studies in Ref. [25] using F344 rats however suggest that after 6 months of inorganic Hg exposure, Hg did not seem to increase appreciably over time.

The rather high retention of Hg in the kidney (and probably other organs), has been attributed to the presence of metallothionein, a low molecular weight protein induced in the kidney following exposure to organic and inorganic Hg [7,38]. This protein acts as a sink for Hg and its binding to the metal has been postulated to protect the kidney against the toxicity of Hg [7,38]. The involvement of an estrogen-dependent mechanism has also been postulated [39].
In contrast to other organs, the brain did not accumulate Hg to any appreciable extent. This is consistent with reports of previous studies in rats and other animals [8,40] indicating that penetration of the lipid architecture of the BBB by the ionic Hg is minimal; although it has been suggested that with a prolonged chronic exposure, accompanied by a slow enough elimination, a small percentage of the inorganic Hg form could be reduced back to the elemental form thereby allowing passage into the brain with subsequent toxicity [5,8,10,40].

Another major finding of this study was that sub-chronic exposure to low doses of inorganic Hg was associated with lipotoxic and non-lipotoxic perturbations in (an already established sex-linked) lipid homeostasis and that these toxic responses differ between males and females. Although this sexual dimorphism appears mixed, two broad classes could be identified: circulating-lipid and organ-lipid sexual dimorphisms. The composites of circulating-lipid sexual dimorphism include:

1. Hypercholesterolemia in the male as compared to hypocholesterolemia in the female
2. Increased plasma phospholipid in male at the highest dose of inorganic Hg (lipotoxicity) as compared to no effect in female animals.
3. Depletion of erythrocyte cholesterol and phospholipid in male (non-lipotoxicity) as compared to enrichment of these lipids in the erythrocyte of female animals (lipotoxicity)
4. Increase in VLDL cholesterol in male animals (lipotoxicity) as compared to decrease observed in female animals (non-lipotoxicity).

Six composites of organ-lipid sexual dimorphism were identified:

1. Cardiac cholesterogenesis in females (lipotoxicity) at the highest dose of inorganic Hg as opposed to depletion of cardiac cholesterol in males (non-lipotoxicity).
2. Hepatic cholesterogenesis in males (lipotoxicity) as opposed to no effect in females.
3. Pulmonary cholesterogenesis in males (lipotoxicity) as compared to no effect in females.
4. Depletion of renal cholesterol in females (non-lipotoxicity) as opposed to no effect in males.

Table 2
Cholesterol phospholipid ratios in different compartments of the animals on exposure to inorganic Hg.

| Compartments | Control | 0.5 mg/kg | 1.0 mg/kg | 1.5 mg/kg |
|--------------|---------|-----------|-----------|-----------|
|               | Male | Female | Male | Female | Male | Female | Male | Female |
| Plasma        | 0.71 ± 0.03a | 0.58 ± 0.04a | 0.77 ± 0.08a | 0.47 ± 0.02b | 0.8 ± 0.04a | 0.26 ± 0.01c | 0.63 ± 0.02b | 0.34 ± 0.03b |
| Erythrocyte   | 0.97 ± 0.05a | 1.80 ± 0.21b | 0.88 ± 0.06a | 1.49 ± 0.16b | 0.68 ± 0.10a | 1.19 ± 0.06b | 1.59 ± 0.15b | 1.32 ± 0.08c |
| HDL           | 0.70 ± 0.08a | 0.34 ± 0.04a | 0.54 ± 0.01a | 0.32 ± 0.03c | 0.61 ± 0.02a | 0.22 ± 0.01b | 0.68 ± 0.04a | 0.31 ± 0.03a |
| Brain         | 0.41 ± 0.02a | 0.20 ± 0.02b | 0.55 ± 0.05a | 0.57 ± 0.02a | 0.38 ± 0.02a | 0.16 ± 0.01b | 0.28 ± 0.02b | 0.18 ± 0.01a |
| Liver         | 1.81 ± 0.07a | 0.60 ± 0.05a | 1.74 ± 0.11a | 0.46 ± 0.03a | 1.77 ± 0.14a | 0.25 ± 0.02b | 2.98 ± 0.27c | 0.16 ± 0.01a |
| Kidney        | 0.25 ± 0.01a | 0.22 ± 0.02a | 0.24 ± 0.01a | 0.20 ± 0.02a | 0.26 ± 0.01a | 0.21 ± 0.01a | 0.30 ± 0.03a | 0.14 ± 0.09a |
| Lung          | 0.20 ± 0.03a | 0.17 ± 0.01a | 0.19 ± 0.01a | 0.20 ± 0.02a | 0.16 ± 0.01a | 0.25 ± 0.02a | 0.16 ± 0.00b | 0.88 ± 0.13a |
| Spleen        | 0.12 ± 0.01a | 0.18 ± 0.02b | 0.13 ± 0.01a | 0.10 ± 0.01a | 0.09 ± 0.02a | 0.17 ± 0.02b | 0.16 ± 0.02a | 0.16 ± 0.02a |
| Heart         | 0.37 ± 0.03a | 0.13 ± 0.01a | 0.20 ± 0.03b | 0.21 ± 0.02a | 0.17 ± 0.02b | 0.25 ± 0.01b | 0.14 ± 0.01b | 0.38 ± 0.01a |

Each value represents the mean ± S.E.M. of 8 rats. Values within a row with different alphabets are significantly different from each other, p < 0.05, compared to control male.

Fig. 10. Effects of inorganic mercury on phospholipid concentrations in brain (A), liver (B), kidney (C), heart (D), lung (E) and spleen (F). Each bar represents the mean ± S.E.M. of 8 rats. Bars with different alphabets are significantly different from each other at p < 0.05 compared to their control counterpart.
5. Pulmonary triglyceride constipation in females (lipotoxicity) as opposed to no effect in males
6. Pulmonary phospholipidosi in male (lipotoxicity) as compared to depletion in females (non-lipotoxicity).

Comorbidly in both sexes, plasma and erythrocyte FFA, hepatic and VLDL phospholipid, brain and erythrocyte triglyceride all increased (lipotoxicity), whereas EM phospholipid and triglyceride, brain and hepatic phospholipid as well as VLDL triglyceride decreased (non-lipotoxicity). Even in these comorbidities, sexual dimorphism was still observed as the concentrations of some lipids in the organ lipid pools were higher in one sex than the other. Although the kinetic mechanisms underlying this sexual dimorphism are not understood, several considerations could be made concerning these observations.

Utilisation of major fuel nutrients such as carbohydrate and lipid to meet both structural and energy requirements, constitutes the foundation of all life activities. In order to maintain this homeostasis, the organism must regulate the flux of the carbon skeletons arising from these fuel molecules in its diverse metabolic processes. A distortion of this flux could spell serious pathophysiological consequences for the organism. In this study, we observed a gradual increase in plasma and erythrocyte FFA in both male and female animals exposed to inorganic Hg. While this increase was more in the female for plasma, it was more in the male for erythrocytes. This observation implies that inorganic Hg did not modulate the absorption of dietary lipids in both sexes. Rather, post-absorptive cellular lipid kinetics was the target of inorganic Hg toxicosis.

Circulating FFA are derived from the hydrolysis of triglycerides in the adipose tissue, with subsequent release into the plasma [41] which then equilibrates with the erythrocytes [42]. This hydrolysis is catalysed by adipose triglyceride lipase [43]. Generally, the release of FFA into the blood is stimulated to serve two purposes:

1. Meeting increased tissue demand for the synthesis of cholesterol, triglycerides, phospholipids, cholesterol esters, etc. for use either as structural cellular machinery or signalling molecules.
2. Covering energy requirements of various tissues.

Increased concentrations of plasma FFA suggests an imbalance between release and utilisation of plasma FFA. In higher animals, this imbalance in the flux of plasma FFA resulting in the elevation of this lipid is thought to occur secondary to psychological and physiological stress [44]. Recent experimental evidences indicate that exposure to inorganic Hg induces oxidative stress in exposed animals [8]. Thus, the observation of elevated plasma FFA in the inorganic Hg-exposed animals suggests a Hg-induced activation of adipose triglyceride lipase with the resultant increase in triglyceride hydrolysis, culminating in increased mobilisation of the FFA into the plasma which then equilibrates with that of the erythrocytes. Consistent with previous studies in humans [41], our findings (Fig. 7) indicate that lipolysis (rate of appearance of FFA in bloodstream) in rats is also sexually dimorphic. While this rate was greater in female animals than males for the plasma, the rate of equilibration of plasma FFA with erythrocyte FFA was greater in males than their female counterparts and that both phenomena were further perpetuated by inorganic Hg exposure. Further consequences of this sexually-induced pathological flux of circulating FFA in the animals can also be considered.

The distribution of circulating lipids between organs is mediated at the level of lipoprotein lipase (LPL), an enzyme located on the walls of blood capillaries of many tissues [45]. Expressed per cell, females have a higher LPL activity than males [46]. LPL catalyses the hydrolysis of plasma triglyceride carried in triglyceride-rich lipoproteins, chylomicrons and VLDL [47], yielding FFA and glycerol in the process. This hydrolysis determines the influx of FFA into tissues and the amount to be oxidised for fuel or esterified for storage, thus making LPL the node for the entry and re-esterification of FFA in those tissues [48,49]. Accumulation of FFA (as observed in this study) is known to inhibit LPL [50]. The downstream effects of LPL inhibition include severe hypertriglyceridemia (secondary to underutilisation of chylomicron and VLDL), hypercholesterolemia and decrease in HDL cholesterol and phospholipids [51]. Data in Figs. 2 and 3 however indicate that while hypercholesterolemia was observed in female animals on exposure to inorganic Hg, plasma of male animals responded with hypercholesterolemia. Furthermore, while hypertriglyceridemia and decrease in HDL cholesterol were observed in both sexes, HDL phospholipid decreased in male animals, but increased only in female animals exposed to 1.5 mg Hg/kg body weight. Although the activity of LPL was not measured in this study, the “discrepancies” observed in the downstream effects of LPL inhibition between male and female animals on exposure to inorganic Hg suggest that LPL inhibition alone cannot account for the observed effects of inorganic Hg. In view of the dyslipidemia observed in HDL (Fig. 3), the involvement of hepatic lipase cannot be ruled out, since this enzyme is known to hydrolyse triglycerides and phospholipids in the HDL, making the HDL remnant smaller, less dense, and hence pro-atherogenic [52]. Although it remains plausible that Hg could bind hepatic lipase to accentuate this sexual dimorphism in the downstream effects of LPL inhibition, experimental evidence in support of this proposition is yet lacking.

Uptake of FFA by tissues is normally stimulated by the signal transduced when there is a high concentration of circulating FFA [53, 54]. Uptake could be either by passive diffusion or via a protein-carrier-mediated pathway [55,56]. Protein carriers that have been characterised include fatty acid translocase (FAT/CD 36), plasma membrane fatty acid binding protein and fatty acid transport protein [55,56].

Following uptake, the FFA are incorporated very rapidly into various lipid pools (phospholipids, triglycerides, cholesterol, cholesterol esters and diglycerides) with a portion serving as substrate for β-oxidation [47]. Data in Figs. 8 to10 indicate that inorganic Hg exposure not only modulated the uptake of FFA by the organs in a sexually dimorphic manner, the lipid pools arising from the utilisation of the FFA, as well as the intra-organ lipid kinetics, were also modulated by inorganic Hg exposure and this modulation also exhibited sexual dimorphism. While inorganic Hg did not inhibit the hepatic uptake of FFA in both sexes, the lipid pools arising from utilisation of FFA in each of the sexes differed. While the liver of male animals incorporated the FFA into the major lipids (cholesterol, triglyceride and phospholipid), their female counterparts incorporated the FFA into triglyceride and phospholipid with no evidence of cholesterol synthesis as female hepatic cholesterol remain unchanged. Furthermore, while there was a preponderance of cholesterol in the male hepatic lipid pool, a high concentration of phospholipid characterised the hepatic lipid pool of female animals exposed to inorganic Hg. There was no evidence of renal uptake of FFA by the female animals. Rather, exposure to inorganic Hg depleted the female kidney of its available triglyceride, suggesting that this lipid was hydrolysed to yield FFA which was then directed to β-oxidation. In contrast however, the slight uptake of FFA observed in male kidney was incorporated into triglyceride.

Pulmonary sexual dimorphism in lipid pool was characterised by cholesterogenesis and phospholipidosis in the male as compared to only pulmonary triglyceride constipation and depletion of phospholipid in the female lipid pool. Inorganic Hg did not inhibit cardiac uptake of FFA in both sexes as evidenced by the characteristics of the lipid pools in this organ – cholesterogenesis and triglyceride constipation in the female, whereas only triglyceride increased in the male. Evidence in support of a slight splenic FFA uptake by the female was provided by the transient phospholipidosis observed in the female spleen on exposure to 0.5 mg Hg/kg body weight whereas inorganic Hg inhibited FFA uptake by the male spleen as evidenced by the depletion of the lipid pool of this organ.

Coming after adipose tissue, the brain epitomises a tissue in mammals with both the highest lipid content and the highest diversity in the composition of these lipids [57]. Unlike the adipose tissue which largely stores FFA as triglycerides for subsequent utilisation and mobilisation to
other peripheral tissues, the brain is thought to mainly utilise acylated lipids to generate phospholipids which are then incorporated into cell membrane elements [58]. Contrary to previously held theories, a number of studies has shown that FAA are able to cross the BBB and be transported into the brain from the systemic circulation [59,60]. Although the brain did not accumulate Hg to any appreciable extent in this study, the observation of depletion of brain cholesterol and phospholipid as a result of Hg exposure is consistent with alteration of the integrity of BBB via increased membrane permeability with subsequent leakage of these lipids from the brain [61]. Similar observation was also made by Teixeira et al. [8] using even a lower dose of inorganic Hg. The BBB is rich in plasmalogens and the binding of Hg to these lipids with high affinity has been shown to alter their biophysical behaviour resulting in breakdown of their aggregates [61,62].

One of the first targets of Hg upon absorption from the gastrointestinal tract are the red blood cells [61]. Lipidomic studies of these cells indicate that their membranes are also rich in plasmalogens [63,64]. A similar alteration in their physico-chemical properties resulting in increased membrane permeability may also be assumed for the depletion of their lipids as a result of inorganic Hg exposure [62].

Although previous studies in humans showed that non-oxidative FAA clearance through re-esterification is higher in females than in males [9,65], and thus suggesting that female tend to store whereas male tend to oxidise circulating FFA [65,66], the characteristics as well as the kinetics of organ lipid pools in this study provide ample evidence that in both sexes, all the non-oxidative clearance of circulating FAA (reflected as triglyceride constipation and phospholipidosis), was induced by inorganic Hg exposure. These are hallmarks of inorganic Hg-induced lipotoxicity.

In order to prevent lipotoxicity, the hepatic tissue in each organism produces VLDL, incorporates into this lipoprotein the excess lipids and then secretes this lipid-rich VLDL into circulation [46,47]. Data in Fig. 4 indicate that while Hg did not inhibit the production of VLDL, the production of this lipoprotein did not keep pace with the increasing influx and esterification of FAA as evidenced by the altered lipid levels of the VLDL, thus allowing cholesterol, triglyceride and phospholipids to accumulate in the liver of male animals, but only triglyceride and phospholipids to accumulate in the liver of female animals which in turn resulted in fatty liver. Although in normal physiology the liver coordinates the metabolism of lipids to meet metabolic requirements, the molecular aspects of this biology are regulated by sex hormones (estrogens and androgens) in both male and female [46]. To what extent Hg influences the sex hormones (or vice versa) to modulate the molecular aspects of this hepatic lipid biology whose effects then reverberate in extra-hepatic tissues as observed in this study, remains to be elucidated.

The molar ratio of cholesterol to phospholipids is one of the indices of membrane fluidity [67,68]. An increase in this ratio indicates decreased fluidity while a decrease indicates increased fluidity [68,69]. Neither of this is desirable for the cell. However, Hg decreased this ratio in the brain of both sexes, decreased it in male heart, but increased it in female heart. Both increases and decreases were also observed in other organs. Several studies have associated changes in lipid/phospholipid profiles with impaired signal as well as energy transductions in heavy metal exposure [70,71]. Compelling accumulated and emerging experimental data from animal and human studies have shown a positive association between high circulating FAA and pathologies such as diabetes mellitus, thyrotoxicosis cardiovascular and pulmonary diseases [44,72–75]. Elevated pulmonary cholesterol has been shown to inhibit surfactant function by disrupting the physiology and turnover of surfactant, leading to impairment in lung mechanics [76]. Abnormalities in surfactant metabolism are the leading causes of acute respiratory distress syndrome, acute lung injury and a diverse array of other respiratory diseases [77]. Even though the biological roles of phospholipids are still not fully understood, depletion of these lipids in nervous tissue is associated with various neurodegenerative disorders including Alzheimer’s disease and Down syndrome [78]. In Alzheimer’s disease, phospholipids are specifically depleted in the areas of the brain showing degeneration and the prognosis of the disease is proportional to this phospholipid depletion [79]. Since Hg has been shown to play a role in Alzheimer’s disease [80] and other pathologies [7], the observed lipotoxic and non-lipotoxic dyslipidemia in this study might be one of the incipient pathophysiological mechanisms underlying Hg-induced pathologies.

In conclusion, the findings of this study indicate that sexual dimorphism was the hallmark of the toxicokinetics and toxicodynamics of sub-chronic inorganic Hg exposure in the rat. While sexual dimorphism in Hg toxicokinetics was characterised by higher levels of Hg in female organs compared to males, the attendant toxicodynamics of Hg exposure reflected as lipotoxic and non-lipotoxic dyslipidemia, exhibited different patterns in both sexes. Since these different responses were observed with Hg concentrations that are environmentally relevant, our findings underscore the need to re-examine therapies directed at Hg as well as other metal intoxication in order to reflect both sexes and remove the currently held bias against the female sex. This may lead to developing novel sex-specific tailored treatment approaches to ameliorate the long-term metabolic risks associated with exposure to Hg and other metals in the general population.

Declaration of competing interest

The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

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References

[1] M. Gochfeld, Sex differences in human and animal toxicology, Toxicol. Pathol. 45 (2017) 172–189.
[2] D.J. Clegg, F. Marvais-Jarvis, An integrated view of sex differences in metabolic physiology and disease, Mol. Metab. 15 (2018) 1.
[3] R.M. Craft, Sex differences in drug- and non-drug-induced analgesia, Life Sci. 72 (2003) 2675–2686.
[4] M. Kogevisan, S.H. Zahm, Introduction: epidemiologic research on occupational health in women, Am. J. Ind. Med. 44 (2003) 563.
[5] J.C. Clifton, Mercury exposure and public health, Pediatr. Clin. North Am 54 (2007) 237–269.
[6] C.C. Bridges, R.K. Zalups, Transport of inorganic mercury and methylmercury in target tissues and organs, J. Toxicol. Environ. Health B Crit. Rev. 13 (2010) 385–410.
[7] J.D. Park, W. Zheng, Human exposure and health effects of inorganic and elemental mercury, J Prev Med Public Health 45 (2012) 344–352.
[8] F.B. Teixeira, A.C. de Oliveira, L.K. Leito, N.C. Fagundes, R.M. Fernandes, L.M. Fernandes, M.C. da Silva, L.L. Amado, F.E. Sagica, E.H. de Oliveira, Exposure to inorganic mercury causes oxidative stress, cell death, and functional deficits in the motor cortex, Front. Mol. Neurosci. 11 (2018) 125.
[9] J.B. Nielsen, O. Andersen, Transplacental passage and fetal deposition of mercury after low-level exposure to methylmercury–effect of seleno-L-methionine, J. Trace Elem. Electrolytes Health Dis. 6 (1992) 227–232.
[10] L.A. Broumand, C.A. Hammett-Stabler, R.E. Winecker, J.D. Ropero-Miller, The toxicology of mercury, Lab. Med. 33 (2002) 614–625.
[11] S. Doering, S. Bose-O’Reilly, U. Berger, Essential indicators identifying chronic inorganic mercury intoxication: pooled analysis across multiple cross-sectional studies, FlexiOne 11 (2016), e0150322.
[12] C.A. Doyyama, A. Neri, Y.-C. Lo, J. Durant, P.J. Dargan, T. Svarthout, O. Biya, S. O. Gidado, S. Haladu, N. Sani-Gwarzo, Outbreak of fatal childhood lead poisoning related to artisanal gold mining in northwestern Nigeria, Environ. Health Perspect. 120 (2012) (2010) 601–607.
[13] D. Joshi, D.K. Mittal, S. Shukla, A.K. Srivastava, S.K. Srivastava, N-acetyl cysteine and selenium protects mercurlic chloride-induced oxidative stress and antioxidant defense system in liver and kidney of rats: a histopathological approach, J. Trace Elem. Med. Biol. 28 (2014) 218–226.
[14] M.S. Trebucovich, M.H. Hazehoff, A.A. Chevalier, S. Passamonti, A. Brandoni, A. M. Torres, Protein expression of kidney and liver bilitranslocase in rats exposed to mercuric chloride—a potential biomarker of toxicity, Toxicol. Lett. 225 (2014) 305–310.
[76] B.R. Roszell, J.-Q. Tao, J.Y. Kevin, L. Gao, S. Huang, Y. Ning, S.I. Feinstein, C. H. Vite, S.R. Bates, Pulmonary abnormalities in animal models due to Niemann-Pick type C1 (NPC1) or C2 (NPC2) disease, PLoS One 8 (2013), e67084.

[77] M. Agassandian, R.K. Mallampalli, Surfactant phospholipid metabolism, Biochim. Biophys. Acta Mol. Cell Biol. Lipids 1831 (2013) 612–625.

[78] P. Brites, H.R. Waterham, R.J. Wanders, Functions and biosynthesis of plasmalogens in health and disease, Biochim. Biophys. Acta Mol. Cell Biol. Lipids 1636 (2004) 219–231.

[79] X. Han, D.M. Holtzman, D.W. McKeel Jr., Plasmalogens deficiency in early Alzheimer’s disease subjects and in animal models: molecular characterization using electrospray ionization mass spectrometry, J. Neurochem. 77 (2001) 1168–1180.

[80] J. Mutter, A. Curth, J. Naumann, R. Deth, H. Walach, Does inorganic mercury play a role in Alzheimer’s disease? A systematic review and an integrated molecular mechanism, J. Alzheimer. Dis. 22 (2010) 357–374.