Article
Evaluation of the Toxicity of Chemical and Biogenic Insecticides to Three Outbreaking Insects in Desert Steppes of Northern China

Wenbing Zhang 1, Hao Ren 1, Feilong Sun 1, Tingting Shen 2,3, Shuai Yuan 2,3, Xiwu Gao 4 and Yao Tan 1,3,*

1 College of Horticulture and Plant Protection, Inner Mongolia Agricultural University, Hohhot 010011, China
2 College of Grassland, Resources and Environment, Inner Mongolia Agricultural University, Hohhot 010010, China
3 Key Laboratory of Grassland Resources, Ministry of Education, Hohhot 010010, China
4 Department of Entomology, China Agricultural University, Beijing 100193, China
* Correspondence: 850310.tanhuaf4@imau.edu.cn; Tel.: +86-157-3471-5085

Abstract: The locusts Oedales asiaticus (Bey-Bienko) and Myrmeleotettix pulpalis (Zubovski) (Orthoptera Acrididae) and the leaf beetle Galeruca daurica (Joannis) (Coleoptera, Chrysomelidae) are economically devastating insect species in the desert steppes of Northern China. Control is mainly and frequently dependent on highly toxic chemicals. To date, there have been no complete and comprehensive reports of insecticide applications to these key pests. In this study, laboratory bioassays were carried out to determine and compare the toxicity of twelve insecticides to three outbreaking insects, O. asiaticus, M. pulpalis, and G. daurica, from three typical desert steppe regions, SZWQ, XHQ and WLTQQ, respectively. The responses of the two locust species and the leaf beetle were evaluated by topical application and leaf dip bioassay techniques across a range of concentrations to develop dosage–mortality regressions. The insecticides tested included six chemical insecticides (β-cypermethrin, imidacloprid, phoxim, λ-cyhalothrin, methomyl, chlorantraniliprole) and six biogenic insecticides (spinosad, avermectin, rotenone, matrine, azadiractin, and methoxyfenozide). The results showed that phoxim, λ-cyhalothrin, β-cypermethrin and spinosad showed highly toxic activity to O. asiaticus, M. pulpalis, and G. daurica, while methonyl, chlorantraniliprole, and rotenone were moderately toxic to both locust species and the leaf beetle. The LC50 values of matrine, azadiractin, and avermectin were more than 1 μg a.i./adult for O. asiaticus and M. pulpalis, the LC50 values of which were higher 2 g/L for G. daurica. Our findings complement information from previous similar studies and will inform future studies relating to the control of outbreaking insects, such as O. asiaticus, M. pulpalis, and G. daurica in desert steppes of northern China. This study is also expected to provide basic data on the use of chemical and biogenic insecticides for application in desert steppes.

Keywords: insecticide toxicity; leaf beetle; locust/grasshopper; bioassay; pest management

Key Contribution: The study reported the susceptibilities of three key pest species to six chemical insecticides and six bio-insecticides in three typical desert steppe sites in northern China and provides a theoretical basis for pest management strategies for steppe and grassland protection.

1. Introduction

In China, steppes account for 80% of vegetation and have high biodiversity and conservation value [1]. The northern steppes, accounting for 25% of the national grassland area, play important roles in protecting the ecological environment, developing the socio-economic growth of the regions, and conserving species diversity. However, a large area of steppes is being gradually degraded into desert due to increasing human activity [2]. Grass production has decreased by 30–50% since the 1950s and steppes have been turned into barren land and desert in some severely degraded areas [3]. Along with the reduced productivity of this grassland, ecosystem functions were significantly impaired, including vegetation cover...
and composition, species richness and diversity, the proportion of fine forage, aboveground and underground biomass, and carbon and nitrogen storage [4,5]. Desert steppes are the most arid grassland ecosystem type, occurring in areas with annual precipitation between 150 and 250 mm and under the influence of continental climatic conditions [6]. The *Allium* plant, *Allium polyrhizum* Turcz. ex Regel and several *Stipa* species are widely distributed in the typical desert steppe [7]. Desert steppe systems are generally sensitive to any external disturbance because of their fragile eco-environment system [8]. In recent decades, it is well known that mass outbreaks of pests have had a serious impact on desert steppes, and these outbreaks have been the subject of considerable research emphasis in northern China because of their impact on desert steppe ecosystem processes and functioning.

Sudden and large-scale pest occurrences in desert steppes of northern China have been monitored continuously for many years [9,10]. Outbreaks of these pests have had disastrous consequences for the grassland ecology and the local economy as a whole, with strongly negative effects on agriculture and the development of animal husbandry. The outbreak frequency and impact of the most commonly occurring pest species have been studied with the locusts *Oedaleus asiaticus* Bey-Bienko and *Myrmeleotettix palpalis* Zubovski (Orthoptera, Acrididae), and leaf beetle *Galeruca daurica* (Joannis) (Coleoptera, Chrysomelidae) reported as the three dominant insect species in the vast desert steppe areas of northern China [10–13].

The locust *Oedaleus asiaticus* (Orthoptera, Acrididae) is an important grass-feeding insect widely distributed in steppe and adjacent farmland throughout Northern Asian regions such as Russia, Mongolia, and China. It is reported to be particularly abundant in the Inner Mongolia steppe and has caused huge damage to the livestock industry and ecological environment for many years [14,15]. As the primary consumer, *O. asiaticus* has specific food preferences, and this extreme adaptation has dramatically reduced grassland productivity and competed with other native herbivores and farmed animals for food resources [16,17]. The other economically important insect species in farming-pastoral ecotones in arid areas, *M. palpalis*, has also caused great losses to the farming and crop planting industry and serious damage to steppes, impeding the development of animal husbandry [18]. Earlier studies have reported the locust species involved, their regional distribution [19], biological characteristics [20,21], feeding habits [22], trophic niches and resource utilization [23,24]. The two locust species occupy the same spatial niches and prefer short grass steppes and xerophytic habitats. They have similar feeding habits, mainly on plants of Poaceae and Compositae. The leaf beetle, *G. daurica*, feeds exclusively on *Allium* species (Liliaceae) in the desert steppes of the Inner Mongolia grasslands in China [25]. Sudden outbreaks of this pest were initially reported to have caused great losses to pasture in 2009 and damage has increased extensively year by year [26–28]. These three key insects have one generation annually and over winter lay eggs, which hatch in spring when the pastures turn green and grow.

The application of chemical insecticides with high efficiency, high selectivity, low toxicity and low residue to control insects in desert steppes has the advantages of a good insecticidal effect with rapid results, and is environmentally friendly [29,30]. At present, it is still necessary and effective to control insects when large-area and high-density outbreaks occur on the steppes [31], so chemical control agents should be chosen scientifically and used rationally according to the actual situation of pest density, species, local vegetation, and environmental conditions to maximize the advantages of chemical control and minimize its side effects [32]. Currently, the use of conventional insecticides (organophosphates, carbamates, pyrethroids, and neonicotinoids) remains the most commonly used control tactic in this grassland ecosystem [15,33].

The insecticides used to control grassland insects in this region are often applied as foliar application from the air [34], but it has been recognized that the misuse of chemicals was seriously disrupting the ecosystem’s function, polluting the environment, and inducing pest resurgence and resistance [35]. Highly effective and environmentally safe control approaches have been explored, and new measures and techniques have been put into use.
for pest management [36]. When compared with chemical insecticides, biogenic insecticides (such as microsporidia microbial metabolites, avermectin, and spinosad) and biological insecticides (such as azadirachtin and matrine) are a better control strategy because they have strong selectivity and are safe for humans and livestock.

Outbreaks of *O. asiaticus*, *M. palpalis* and *G. daurica* have caused remarkable grass yield losses and desert steppe deterioration in recent years, thereby putting tremendous pressures on this fragile ecosystem. According to the latest statistics from the China Pesticide Information Official Network, there are 35 pesticide products registered by the Ministry of Agriculture and Rural Affairs of China for grass locust control in China. Among them, the seven main chemical insecticide ingredients are dichlorvos, malathion, imidacloprid, β-cypermethrin, cypermethrin, avermectin and triazophos, and two biogenic insecticides of plant origin are matrine and azadirachtin. More registered insecticides are compounded and although many broad-spectrum insecticides with different modes of action were applied against the main insect species in this region, few studies about current insecticide toxicities and the resistances of the grass pests have been reported. The study reported here documented the susceptibilities of these three key pest species to six chemical insecticides and six bio-insecticides from 2021 to 2022 using both leaf-dip and topical applications in three typical desert steppe sites in northern China, which provides a theoretical basis for pest management strategies for steppe and grassland protection.

2. Results

2.1. Toxicity of Insecticides to *Oedaleus asiaticus* from the Three Regions Studied

*O. asiaticus* adults from SZWQ were most susceptible to the traditionally-used insecticides β-cypermethrin, imidaclorpid, phoxim, and λ-cyhalothrin. The LC$_{50}$ values of the eleven insecticides ranged from 213.48 ng a.i./adult (β-cypermethrin) to 11.99 µg a.i./adult (avermectin), a 56.16-fold range, compared with a much larger range in LC$_{50}$ values for the other two regions. The toxicities of the eleven tested insecticides against *O. asiaticus* SZWQ populations, ranked from high to low (based on LC$_{50}$ values), are β-cypermethrin > imidaclorpid > phoxim > λ-cyhalothrin > spinosad > rotenone > methomyl > chlorantraniliprole > matrine > azadirachtin > avermectin. However, it was found that *O. asiaticus* collected from SZWQ was less susceptible to one chemical insecticide, chlorantraniliprole, as well as three biogenic insecticides, matrine and azadirachtin, and avermectin (Table 1).

The large range of toxicities of all the tested insecticides (avermectin LC$_{50}$/β-cypermethrin LC$_{50}$ = 52.69) to *O. asiaticus* adults from XHQ was remarkable, along with the toxicity range of several biogenic insecticides (avermectin LC$_{50}$/spinosad LC$_{50}$ = 26.41), although the ranking of the eleven tested insecticides from high to low toxicity was similar with that observed with *O. asiaticus* adults from SZWQ. For the *O. asiaticus* population collected from XHQ, the toxicity order was β-cypermethrin > phoxim > spinosad > imidaclorpid > λ-cyhalothrin > rotenone > methomyl > chlorantraniliprole > matrine > azadirachtin > avermectin. Among the biogenic insecticides, avermectin was far less toxic than any of the other biocides, and the assay time was the longest (72 h). Based on LC$_{50}$ values, *O. asiaticus* adults from XHQ were 1.17 and 1.24 times less susceptible to spinosad and rotenone, respectively, when compared with *O. asiaticus* adults from SZWQ, and relative toxicity indexes were 0.67 and 0.66 against matrine and azadirachtin.

The ranking of insecticide toxicity from high to low for the *O. asiaticus* adults collected from WLTQQ in 2022 followed a similar general pattern to the populations from SZWQ and XHQ, which was β-cypermethrin > phoxim > chlorantraniliprole > imidaclorpid > λ-cyhalothrin > spinosad > rotenone > methomyl > azadirachtin > matrine. β-cypermethrin had the highest toxicity towards all *O. asiaticus* populations from SZWQ, XHQ, and WLTQQ when compared to the other test insecticides, indicating that it still retains good insecticidal activity in pest management in desert steppes. There was also a significant difference for chlorantraniliprole toxicity towards *O. asiaticus* between the WLTQQ, SZWQ, and XHQ populations (Table 1) while all the other biogenic insecticides except for spinosad showed low toxicity to *O. asiaticus* adults from the three regions.
Table 1. Toxicity of chemical and biogenic insecticides against three field populations of *Oedaleus asiaticus* Bey-Bienko 2021–2022.

| Insecticide | Chemical Structure | Year | Population | LC₅₀ (95% FL) [ng a.i./Adult] | Slope ± SE | X² (df a) | RTI b (95% FL) |
|-------------|-------------------|------|------------|-----------------------------|------------|----------|-------------|
| Phoxim      | ![Phoxim Structure](image) | 2021 | SZWQ       | 301.58 [232.18–374.85]      | 2.50 ± 0.37 | 0.41 (3) | 1           |
|             |                   |      | XHQ        | 369.96 [289.96–460.10]      | 2.42 ± 0.37 | 0.25 (3) | 1.23         |
|             |                   | 2022 | WLTQQ      | 281.90 [214.42–348.85]      | 2.89 ± 0.46 | 0.65 (3) | 0.93         |
| Methomyl    | ![Methomyl Structure](image) | 2021 | SZWQ       | 940.13 [798.65–1102.83]     | 4.52 ± 0.82 | 0.74 (3) | 1           |
|             |                   |      | XHQ        | 954.24 [783.35–1163.26]     | 3.26 ± 0.57 | 0.99 (3) | 1.02         |
|             |                   | 2022 | WLTQQ      | 906.55 [764.05–1061.96]     | 4.80 ± 0.92 | 0.46 (3) | 0.96         |
| Imidacloprid| ![Imidacloprid Structure](image) | 2021 | SZWQ       | 292.89 [67.35–591.22]       | 0.97 ± 0.18 | 1.42 (3) | 1           |
|             |                   |      | XHQ        | 717.91 [531.21–924.09]      | 2.03 ± 0.31 | 0.23 (5) | 2.45         |
|             |                   | 2022 | WLTQQ      | 462.59 [331.11–612.17]      | 1.80 ± 0.29 | 0.85 (3) | 1.58         |
| λ-cyhalothrin| ![λ-cyhalothrin Structure](image) | 2021 | SZWQ       | 334.90 [178.72–533.40]      | 0.99 ± 0.17 | 0.53 (4) | 1           |
|             |                   |      | XHQ        | 878.04 [560.79–1281.50]     | 1.39 ± 0.26 | 0.48 (3) | 2.62         |
|             |                   | 2022 | WLTQQ      | 546.12 [317.84–788.50]      | 1.48 ± 0.28 | 0.66 (4) | 1.63         |
| β-cypermethrin| ![β-cypermethrin Structure](image) | 2021 | SZWQ       | 213.48 [110.20–370.33]      | 1.06 ± 0.23 | 0.57 (4) | 1           |
|             |                   |      | XHQ        | 286.34 [208.80–405.16]      | 2.10 ± 0.49 | 0.65 (4) | 1.68         |
|             |                   | 2022 | WLTQQ      | 174.95 [89.08–281.70]       | 1.37 ± 0.29 | 0.67 (4) | 0.82         |
| Chlorantraniliprole| ![Chlorantraniliprole Structure](image) | 2021 | SZWQ       | 1116.63 [375.05–4949.58]    | 1.01 ± 0.17 | 2.10 (4) | 1           |
|             |                   |      | XHQ        | 1017.52 [605.26–1709.82]    | 1.04 ± 0.21 | 0.85 (4) | 0.91         |
|             |                   | 2022 | WLTQQ      | 435.50 [171.75–827.09]      | 1.04 ± 0.18 | 1.05 (4) | 0.39         |
| Avermectin c| ![Avermectin Structure](image) | 2021 | SZWQ       | 11.99 µg [9.50–15.70]       | 1.99 ± 0.32 | 0.50 (4) | 1           |
|             |                   |      | XHQ        | 15.09 µg [12.10–19.82]      | 2.34 ± 0.42 | 0.05 (4) | 1.26         |
| Spinosad    | ![Spinosad Structure](image) | 2021 | SZWQ       | 488.06 [301.24–730.80]      | 1.44 ± 0.27 | 0.68 (4) | 1           |
|             |                   |      | XHQ        | 571.48 [388.62–892.36]      | 1.20 ± 0.18 | 0.34 (5) | 1.17         |
|             |                   | 2022 | WLTQQ      | 578.99 [386.96–842.43]      | 1.54 ± 0.26 | 0.92 (4) | 1.18         |
| Matrine     | ![Matrine Structure](image) | 2021 | SZWQ       | 1702.98 [455.41–2149.16]    | 3.16 ± 0.85 | 0.38 (3) | 1           |
|             |                   |      | XHQ        | 1147.01 [748.61–1464.61]    | 2.73 ± 0.62 | 0.81 (3) | 0.67         |
|             |                   | 2022 | WLTQQ      | 1727.01 [1129.87–2221.80]   | 2.87 ± 0.80 | 0.34 (3) | 1.01         |
Table 1. Cont.

| Insecticide | Chemical Structure | Year | Population | LC_{50} (95% FL) [ng a.i./Adult] | Slope ± SE | X^2 (df ^a) | RTI ^b (95% FL) |
|-------------|-------------------|------|------------|---------------------------------|------------|-------------|----------------|
| Azadirachtin | ![Azadirachtin Structure](image) | 2022 | SZWQ       | 1811.89 [1179.77–2393.49]       | 2.54 ± 0.75 | 0.51 (3)    | 1              |
|             |                   | 2022 | XHQ        | 1198.20 [451.57–1820.36]        | 2.53 ± 0.52 | 1.10 (3)    | 0.66           |
|             |                   | 2022 | WLTQQ      | 1069.17 [710.83–1412.23]        | 2.50 ± 0.63 | 0.46 (3)    | 0.59           |
| Rotenone    | ![Rotenone Structure](image) | 2022 | SZWQ       | 750.63 [494.29–1015.93]         | 2.08 ± 0.44 | 0.98 (3)    | 1              |
|             |                   | 2022 | XHQ        | 930.94 [552.91–1622.65]         | 1.17 ± 0.28 | 0.62 (3)    | 1.24           |
|             |                   | 2022 | WLTQQ      | 737.61 [420.29–1028.81]         | 2.10 ± 0.52 | 0.73 (3)    | 0.98           |

^a Degree of freedom. ^b Relative toxic indexes. ^c Test time: 72 h.

2.2. Toxicity of Insecticides to Myrmeleotettix palpalis from SZWQ and XHQ

The large range of toxicities of all tested insecticides to *M. palpalis* populations from SZWQ and XHQ (avermectin LC_{50}/spinosad LC_{50} = 61.21 for SZWQ and 60.64 for XHQ) was remarkable (Table 2). LC_{50} values show that both *M. palpalis* populations were most susceptible to spinosad followed by λ-cyhalothrin, imidacloprid and methomyl. *M. palpalis* from these two regions was the least susceptible to avermectin, usually followed by matrine in the SZWQ population, but followed by azadirachtin in the XHQ population. Rotenone and chlorantraniliprole also generally ranked low in toxicity among the eleven insecticides in *M. palpalis* field populations. The high toxicity of spinosad to SZWQ and XHQ populations indicated that a novel biogenic insecticide not used in this region showed highly toxic activity to *M. palpalis*. Meanwhile, *M. palpalis* from SZWQ and XHQ was relatively susceptible to λ-cyhalothrin (LC_{50} values were 70.73 ng a.i./adult and 88.39 ng a.i./adult) and β-cypermethrin (LC_{50} values were 279.73 ng a.i./adult and 215.54 ng a.i./adult), which demonstrated that pyrethroid insecticides generally have a strong application value for *M. palpalis* control.

Table 2. Toxicity of chemical and biogenic insecticides against two field populations of *Myrmeleotettix palpalis* 2021–2022.

| Insecticide | Year | Population | LC_{50} (95% FL) [ng a.i./Adult] | Slope ± SE | X^2(df ^a) | RTI ^b (95% FL) |
|-------------|------|------------|---------------------------------|------------|-------------|----------------|
| Phoxim      | 2021 | SZWQ       | 216.72 [134.98–290.42]          | 2.03 ± 0.40 | 1.05 (6)    | 1              |
|             | 2022 | XHQ        | 247.89 [163.68–309.71]          | 3.67 ± 0.66 | 1.03 (4)    | 1.14           |
| Methomyl    | 2021 | SZWQ       | 200.95 [126.87–263.76]          | 3.30 ± 0.55 | 1.31 (5)    | 1              |
|             | 2022 | XHQ        | 164.64 [121.86–200.89]          | 2.88 ± 0.52 | 0.44 (4)    | 0.82           |
| Imidacloprid| 2021 | SZWQ       | 101.94 [61.44–172.82]           | 1.10 ± 0.21 | 0.78 (3)    | 1              |
|             | 2022 | XHQ        | 109.99 [87.76–133.16]           | 3.39 ± 0.55 | 0.10 (3)    | 1.08           |
| λ-cyhalothrin| 2021 | SZWQ       | 77.73 [55.00–98.52]             | 2.68 ± 0.50 | 0.21 (3)    | 1              |
|             | 2022 | XHQ        | 88.39 [66.06–114.89]            | 2.11 ± 0.32 | 0.22 (3)    | 1.14           |
### Table 2. Cont.

| Insecticide       | Year | Population | LC<sub>50</sub> (95% FL) [ng a.i./Adult] | Slope ± SE | X2(df<sup>a</sup>) | RTI<sup>b</sup> (95% FL) |
|-------------------|------|------------|----------------------------------------|-----------|---------------------|--------------------------|
| β-cypermethrin    | 2021 | SZWQ       | 279.73 [87.10–684.14]                  | 1.44 ± 0.22 | 2.51 (4)           | 1                        |
|                   | 2022 | XHQ        | 215.54 [95.76–376.44]                  | 1.59 ± 0.27 | 1.55 (4)           | 0.77                     |
| Chlorantraniliprole | 2021 | SZWQ       | 278.06 [118.76–447.50]                | 2.58 ± 0.48 | 1.30 (3)           | 1                        |
|                   | 2022 | XHQ        | 339.89 [257.88–424.45]                | 2.23 ± 0.31 | 0.83 (4)           | 1.22                     |
| Avermectin c      | 2021 | SZWQ       | 2.17 µg [1.33–3.16]                   | 1.80 ± 0.27 | 1.28 (5)           | 1                        |
|                   | 2022 | XHQ        | 3.03 µg [2.35–3.84]                   | 1.96 ± 0.34 | 0.35 (4)           | 1.40                     |
| Spinosad          | 2021 | SZWQ       | 35.45 [23.65–48.43]                   | 1.78 ± 0.32 | 0.28 (4)           | 1                        |
|                   | 2022 | XHQ        | 49.96 [32.71–70.96]                   | 1.69 ± 0.33 | 0.10 (4)           | 1.41                     |
| Matrine           | 2021 | SZWQ       | 526.71 [386.95–665.46]                | 2.37 ± 0.38 | 0.42 (5)           | 1                        |
|                   | 2022 | XHQ        | 749.73 [438.05–1249.35]               | 1.16 ± 0.30 | 1.09 (4)           | 1.42                     |
| Azadirachtin      | 2021 | SZWQ       | 924.86 [720.98–1137.62]               | 2.78 ± 0.56 | 0.56 (3)           | 1                        |
|                   | 2022 | XHQ        | 662.80 [404.34–918.39]                | 2.09 ± 0.63 | 0.66 (3)           | 0.72                     |
| Rotenone          | 2022 | SZWQ       | 337.50 [198.57–469.59]                | 1.85 ± 0.40 | 0.77 (4)           | 1                        |
|                   |      | XHQ        | 198.20 [104.75–282.40]                | 1.92 ± 0.40 | 0.60 (5)           | 0.59                     |

<sup>a</sup> Degree of freedom. <sup>b</sup> Relative toxic indexes. <sup>c</sup> Test time: 72 h.

#### 2.3. Toxicity of Insecticides to *Galeruca daurica* from SZWQ and XHQ

*G. daurica* from SZWQ and XHQ were the most susceptible to imidacloprid as the LC<sub>50</sub> values were 0.17 and 0.16 mg/L, respectively. The eleven insecticides can be ranked from high to low toxicity to *G. daurica* from SZWQ (based on LC<sub>50</sub> values) in the order imidacloprid > λ-cyhalothrin > β-cypermethrin > phoxim > methomyl > spinosad > chlorantraniliprole > rotenone > matrine > methoxyfenozide > azadirachtin with overlaps of 95% fiducial limits (Table 3). Ranking insecticide toxicity from high to low for *G. daurica* from XHQ followed the similar general pattern as with *G. daurica* from SZWQ (imidacloprid > λ-cyhalothrin > phoxim > β-cypermethrin > spinosad > methomyl > chlorantraniliprole > rotenone > matrine > azadirachtin > methoxyfenozide). LC<sub>50</sub> values for *G. daurica* from SZWQ and XHQ were 20,335, 18,641, and 18,856, 23,767 times more susceptible to imidacloprid, respectively, when compared with azadirachtin and methoxyfenozide. The high toxicity of imidacloprid to *G. daurica* from SZWQ and XHQ indicated that this neonicotinoid had good application values. Meanwhile, *G. daurica* from SZWQ and XHQ was more susceptible to λ-cyhalothrin (LC<sub>50</sub> values were 0.73 and 0.56 mg/L) and β-cypermethrin (LC<sub>50</sub> values were 1.57 mg/L and 1.06 mg/L). Both the SZWQ and XHQ *G. daurica* populations were least susceptible to azadirachtin and methoxyfenozide, which was usually followed by matrine. Rotenone also generally had low toxicity among the eleven insecticides while...
all the biogenic insecticides except spinosad showed high LC\textsubscript{50} values for \textit{G. daurica} from SZWQ and XHQ.

Table 3. Toxicity of chemical and biogenic insecticides against two field populations of \textit{Galeruca daurica} 2021–2022.

| Insecticide        | Year | Population | LC\textsubscript{50} (95% FL) [mg/L] | Slope ± SE | X\textsuperscript{2}(df \textsuperscript{a}) | RTI \textsuperscript{b} (95% FL) |
|--------------------|------|------------|-------------------------------------|------------|--------------------------------|---------------------------|
| Phoxim             | 2022 | SZWQ       | 2.91 [1.53–4.74]                   | 1.39 ± 0.27 | 0.20 (3)                        | 1                         |
|                    | 2022 | XHQ        | 0.66 [0.26–1.24]                   | 1.09 ± 0.23 | 0.72 (3)                        | 0.26                      |
| Methomyl           | 2022 | SZWQ       | 40.77 [27.30–52.25]                | 2.99 ± 0.80 | 0.54 (3)                        | 1                         |
|                    | 2022 | XHQ        | 31.91 [22.55–39.91]                | 3.12 ± 0.69 | 0.79 (3)                        | 0.78                      |
| Imidacloprid       | 2022 | SZWQ       | 0.17 [0.10–0.31]                   | 0.97 ± 0.15 | 0.59 (3)                        | 1                         |
|                    | 2022 | XHQ        | 0.16 [0.07–0.32]                   | 0.85 ± 0.17 | 0.28 (4)                        | 6.82                      |
| \textalpha-cyhalothrin | 2022  | SZWQ     | 0.73 [0.46–1.09]                   | 1.69 ± 0.28 | 0.66 (3)                        | 1                         |
|                    | 2022  | XHQ     | 0.56 [0.30–0.92]                   | 1.24 ± 0.22 | 0.25 (3)                        | 0.77                      |
| \textbeta-cypermethrin | 2022  | SZWQ     | 1.57 [0.91–2.61]                   | 1.41 ± 0.27 | 0.55 (3)                        | 1                         |
|                    | 2022  | XHQ     | 1.06 [0.55–1.86]                   | 1.16 ± 0.24 | 0.58 (3)                        | 0.68                      |
| Chlorantraniliprole | 2021 | SZWQ     | 56.12 [42.69–71.81]                | 2.03 ± 0.27 | 0.27 (4)                        | 1                         |
|                    | 2021 | XHQ     | 38.17 [12.27–105.19]               | 1.40 ± 0.20 | 2.65 (4)                        | 0.68                      |
| Methoxyfenozide    | 2021 | SZWQ       | 3205.64 [2088.01–4319.27]          | 2.80 ± 0.59 | 1.12 (4)                        | 1                         |
|                    | 2021 | XHQ       | 3802.80 [3008.44–4738.47]          | 3.53 ± 1.19 | 0.04 (3)                        | 1.19                      |
| Spinosad           | 2022 | SZWQ       | 50.10 [33.78–64.40]                | 3.00 ± 0.66 | 0.30 (3)                        | 1                         |
|                    | 2022 | XHQ       | 28.12 [8.64–39.70]                 | 2.46 ± 0.81 | 0.06 (2)                        | 0.56                      |
| Matrine            | 2021 | SZWQ       | 2267.18 [902.28–3977.09]           | 1.78 ± 0.35 | 1.11 (4)                        | 1                         |
|                    | 2021 | XHQ       | 2269.82 [1384.18–3492.35]          | 1.45 ± 0.26 | 0.67 (4)                        | 4.41                      |
| Azadirachtin       | 2021 | SZWQ       | 3456.99 [1991.59–5242.97]          | 1.60 ± 0.42 | 0.20 (3)                        | 1                         |
|                    | 2021 | XHQ       | 2982.66 [1788.12–4133.05]          | 1.91 ± 0.38 | 0.10 (4)                        | 0.86                      |
| Rotenone           | 2021 | SZWQ       | 751.73 [529.19–1050.16]            | 1.91 ± 0.38 | 0.82 (3)                        | 1                         |
|                    | 2021 | XHQ       | 660.39 [350.54–1096.59]            | 2.17 ± 0.34 | 1.18 (3)                        | 1.44                      |

\textsuperscript{a} Degree of freedom. \textsuperscript{b} Relative toxic indexes. \textsuperscript{c} Test time: 240 h. \textsuperscript{*}.

3. Discussion

In recent years, outbreaks of \textit{O. asiaticus}, \textit{M. palpalis}, and \textit{G. daurica} have become frequent in the northern China steppes due to livestock overgrazing, extreme climate change, and consequent land desertification [33,37]. These pest species prefer to live on heavily-grazed steppes and mainly feed on gramineaceous and liliaceous plants such as stipa, wild rye, and allium [16,27,28]. The outbreaks have brought remarkable grass yield losses and led to grassland deterioration, thereby adversely affecting animal husbandry production and the development of the grassland industry. Chemical control is thus far the major means of controlling the outbreaks of insects in northern China [15,38]. Over the years, pyrethroids have been commonly used to control insect outbreaks in northern grasslands.
and steppes, among which β-cypermethrin was sprayed by local grassland stations as required, and used intensively on an annual basis. However, their effectiveness has been reported to have gradually diminished (based on insecticide resistance investigations), most likely because some insect populations in different regions have already developed resistance to these pyrethroids [16,39].

The toxicity evaluation showed that the insecticides β-cypermethrin, imidacloprid, phoxim, and λ-cyhalothrin displayed high toxicity to O. asiaticus by direct contact at the dosage recommended on the desert steppes from the three tested regions. There were great differences in the susceptibilities of O. asiaticus to the various insecticides, mainly due to the different mechanisms of action of each class of insecticide. The chemical insecticides (i.e., pyrethroids, organophosphates, and neonicotinoids) were more efficient in controlling O. asiaticus than the biogenic insecticides (i.e., avermectin, azadirachtin, and matrine) based on the bioassay results under laboratory conditions. Pyrethroids and organophosphates have a strong thixotropic and stomachic effect, and have shown greater efficiency than other insecticides tested when used by topical application, and which may be the consequence of their greater solubility in acetone (>450 g L$^{-1}$ at 20°C), enhancing insect epidermal penetration and diffusion in the tissues [40]. Of the three types of traditionally used chemical insecticides, dichlorvos, triazophos, malathion, beta-cypermethrin, cypermethrin, avermectin, imidacloprid, and their compound products are registered for production, sale and use against locusts in China [41]. Although phoxim and lambda-cyhalothrin are not registered on the Official China Pesticide Information Website, they have good efficiency and might be used to manage the grassland locusts in future. Among them, β-cypermethrin has been the most commonly used insecticide for controlling O. asiaticus in Inner Mongolia over the last few decades, and the spraying dosage used was dependent on the density of O. asiaticus [42]. However, O. asiaticus has been reported to become resistant to pyrethroids with increasing frequency and usage [43]. Dong et. al. (2015) investigated the susceptibility to β-cypermethrin of O. asiaticus populations obtained from multiple areas of the Inner Mongolian steppes, and their results illustrated that the susceptibility of O. asiaticus populations to β-cypermethrin has been continuously reduced, which they attributed to elevated enzyme activities and mRNA expression levels of CarE and GST [16]. Given the emergence of increased insecticide resistance in O. asiaticus to conventional insecticides in field application, the rotation of insecticides with different modes of action to prevent the development of insecticide resistance in conventional systems is strongly suggested. Our results also illustrated that the biogenic insecticides showed high LC$_{50}$ values to O. asiaticus adults from all three regions, probably because biogenic insecticides have shortcomings such as the longer time taken to cause pest mortality, and hence a more unstable control effect [44]. In fact, external biotic or abiotic factors including temperature, humidity, density, and the height of vegetation also significantly affect the control efficacy of biogenic insecticides [45]. At present, the main biological control agents against O. asiaticus include fungal insecticides (such as: Metarhizium species, Beauveria bassiana Bals.), poxvirus, and natural enemies (such as birds and small predators). Among them, M. anisopliae has been proven to have the most effective control effect against O. asiaticus; it has the advantage of being green, safe, and has long duration in the field, but has a slow effect on outbreaking insects [46]. Chemical insecticides are fast-acting and highly effective, but their excessive use generates residues which pollute the environment and can also have a negative impact on food safety for humans and animals [47]. However, the use of a combination of these insecticides can effectively solve these challenges by working synergistically to achieve better pest control, and can also address the negative effects of excessive use of chemical agents [48].

Chemical and biological control are both important for the management of M. palpalis in the desert steppes of northern China [49]. Traditional insecticides such as phoxim, methomyl, λ-cyhalothrin, and β-cypermethrin have been used in these bio-systems for many years, but they may also lead to reductions in insect diversity and natural enemy populations. In the past decade, insecticides with novel and unique modes of action have...
shown high toxicity to insect populations while being relatively non-toxic to natural enemies [50]. Our results suggest that spinosad could prove to be essential to the integration of chemical and biological control in IPM programs, as it was the most toxic insecticide to *M. palpalis* under laboratory conditions. However, we also observed the low mortality of *O. asiaticus* after spinosad topical application, suggesting that the toxicity of spinosad may vary depending on the route of exposure for specific species. The toxicity of spinosad to *M. palpalis* was higher than that of pyrethroids, carbamates, and organophosphates, and it may also provide better selectivity to predators when compared with conventional insecticides if applied in the field [51]. Spinosad is more toxic by ingestion than by contact, so natural enemy insects that do not feed on treated plant tissues can therefore be protected from this insecticide [52], giving it the potential for use in the pest management in the northern steppes. In this study, broad-spectrum insecticides (i.e., λ-cyhalothrin, methomyl, and phoxim) were relatively highly toxic to *M. palpalis*, while biogenic insecticides (i.e., avermectin, azadirachtin, and matrine) were less toxic under laboratory conditions. Biogenic insecticides have, however, received extensive research attention and have broad prospects for development and application due to their positive characteristics, such as low toxicities, easy degradation, miscibility with other pesticides, and ability to control insects that have already developed resistance [53]. Avermectin, azadirachtin, and matrine have already been registered and applied. The results of this study indicate that spinosad is highly toxic and imidacloprid is moderately toxic to *M. palpalis* when compared to conventional insecticides, suggesting that both spinosad and imidacloprid would likely contribute to pest management in desert steppes and should be recommended. Spinosad and imidacloprid were selective with regard to acute toxicity, but further work is needed to evaluate their residual toxicity and their potential sublethal effects.

For the outbreaking leaf beetle, *G. daurica* in northern grassland, λ-cyhalothrin, β-cypermethrin, imidacloprid and phoxim were the most toxic insecticides when tested by a leaf-dip bioassay method. Among them, only β-cypermethrin and imidacloprid are officially registered insecticides for controlling leaf beetles. Exposure to imidacloprid may cause the reduced motility of *G. daurica*, probably because of the interaction of imidacloprid with acetylcholine receptors, resulting in a modification of receptor conformation [54]. There was an exception for spinosad when ingested, which caused high mortality. Our results can therefore be used as guidelines regarding which insecticides and spraying methods may be toxic to key insect species, and this information may assist pastoralists or farmers during insecticide spraying times. For several biogenic insecticides, matrine, azadirachtin, and methoxyfenozide all showed low toxicity to *G. daurica* larvae from two regions, probably because biogenic insecticides require a long time to produce a toxic effect [44]. Among them, methoxyfenozide was reported as having low acute toxicity to *G. daurica*, when it was first applied in the desert steppes of northern China. Methoxyfenozide induces toxic symptoms in insects identical to those produced by other diacylhydrazine ecdysone agonists: it inhibits feeding, induces premature apolysis, causes abnormal cuticle deposition and other molting irregularities, and inhibits ecdysis in larval insects [55–57]. Therefore, the use of a combination of methoxyfenozide and other chemical insecticides can effectively achieve better pest control of *G. daurica* by producing a synergistic efficacy.

It is essential that insecticides with different modes of action are effective and long-lasting under field conditions while these grassland ecosystems are regenerating. The rational use of combinations of these chemical and biogenic insecticides may delay the development of insecticide resistance and prolong the time post-spraying that products are effective; an extended time of effectiveness would increase the possible exposure of insects to the active ingredients of insecticides, which are highly recommended for development as an important chemical control strategy for pest management in the desert steppes of northern China. Our data provide valuable information to develop the rational use of chemical and biogenic insecticides in a framework of integrated pest management in the desert steppes of northern China and to recommend some potential insecticide application strategies as follows: (i) officially registered insecticides are strongly recommended to...
control target key insects in desert steppes and grasslands. (ii) Pesticides of the same type with good efficiency can be used in rotation, such as: phoxim, methomyl, lambda-cyhalothrin etc. (iii) The use of combinations of chemical and biogenic insecticides with different modes of action can be adopted to expand the insecticidal spectrum and delay resistance. (iv) Insecticides with novel and unique modes of action should be investigated to evaluate their toxicity in labs and under field conditions. (v) Insecticide resistance levels should be monitored constantly. As this work was carried out under laboratory conditions, further studies will be carried out under field conditions to determine the best formulations of compounds which show promising efficacy.

4. Conclusions

The use of Oedaleus asiaticus, Myrmeleotettix palpalis, and Galeruca daurica in toxicity tests in the present study confirmed the insecticidal effects of various types of chemical and biogenic insecticides. The results of this study will complement information from previous similar studies and inform future studies related to the control of outbreaking insects, such as O. asiaticus, M. palpalis, and G. daurica in the desert steppes of northern China. This research also provided baseline data on the use of chemical and biogenic insecticides for application in desert steppes.

5. Materials and Methods

5.1. Insects

The field populations of three insect species, O. asiaticus, M. palpalis, and G. daurica used for insecticide toxicity evaluation, were collected from three locations in typical desert grasslands of northern China from 2021 to 2022. The outbreaks of the two locust species occur in late May and last until early July, while populations of the leaf bug G. daurica have high density during May. The collection sites are located in Xianghuang Qi (Xilinhot, Inner Mongolia, China. 40°98′ N, 113°54′ E, XHQ), Siziwang Banner (Agro-pastoral Field Experiment station, Ulanqab, Inner Mongolia, China. 41°22′ N, 111°21′ E, SZWQ), and Urat Qianqi (Bayannur, Inner Mongolia, China. 40°98′ N, 108°91′ E, WLTQQ) (Figure 1). The fifth-instar nymphs of O. asiaticus and M. palpalis were collected using sweep nets in late May and early June of 2021 and 2022 and immediately processed for bioassays.

![Figure 1](image-url)  
**Figure 1.** Location of the three sampling sites in desert grasslands of northern China used for evaluating toxicity in three outbreaking insects.
5.2. Insecticides and Chemicals

The active insecticide ingredients phoxim (99%) were obtained from TianJin Pesticide Co., Ltd. (Tianjin, China); methomyl (90%) was from Jiangsu Fengshan Group Co., Ltd. (Yancheng, China); lambda-cyhalothrin (98%) and imidacloprid (95%) were purchased from Jiangsu Changlong Chemical Co., Ltd. (Changzhou, China); avermectin (91.2%) and beta-cypermethrin (95.2%) were gifted by HeBei VEYong Bio-chemical Co., Ltd. (Shijiazhuang, China) and Tianjin Longdeng Chemical Co., Ltd. (Tianjin, China), respectively. Methoxyfenozide was from Jiangsu Agricultural Academy. Technical grade 95% chlorantraniliprole was obtained from DuPont Agricultural Chemicals Ltd. (Shanghai, China).

The active biogenic insecticides azadirachtin (30%), and matrine (98%) were provided by Beijing Qingyuanbao Biological Technology Co., Ltd. (Beijing, China), and the analytical grade rotenone (96%) was brought online from SIGMA Co., Ltd. The other chemicals and solvents were purchased from commercial suppliers, including acetone, ethyl alcohol, ethyl acetate, and CO\textsubscript{2} gas.

5.3. Bioassay

5.3.1. Topical Application

The toxicity of each insecticide was determined by topical application. Groups of 30 fifth-instar nymphs were anaesthetized using *Drosophila* anesthesia equipment (YH-DACO2, Yi Hong technology company, Wuhan, China), and a 2.5 µL (*O. asiaticus*), 1 µL (*M. palpalis*) droplet of technical-grade insecticide in acetone was administered to the third abdominal segment using a semi-automatic dropper (PB-600 PAT, 3161323, Hamilton Company of America, Nevada, USA). Each insecticide was dissolved in acetone as a stock solution and serially diluted 5–8 times in a gradient, resulting in 10–90% mortality of the test insects. The control fifth-instar nymphs were treated with a 1% acetone solution. Each treatment was replicated three times with each repetition containing 10 fifth nymphs per plastic box (10 cm deep, 15 cm in diameter at top and bottom), meaning a total of 30 fifth-instar nymphs were used to test each of the insecticide or control, and all assays were performed at 22 °C ± 2 °C, RH 50% ± 10% and L 14 h: D 10 h. Nymphs that did not show any coordinated movement after probing with a soft brush were assumed to be dead after 48 h exposure.

5.3.2. Leaf-Dip Bioassay

The toxicity of insecticides on *G. daurica* was determined using the leaf-dip bioassay method as described by Shelton (1993) with minor modifications. Stock solutions of each of the insecticides were prepared in 0.05% Triton X-100 sterile water to 5–8 dilutions, causing 10–90% mortality. The leek host plants, *Allium hookeri*, were planted in a greenhouse, grown without any pest-control chemicals or chemical fertilizers. Fresh leaves were cut into 2.5 cm diameter pieces and for each replicate, five leaf pieces were dipped in a solution of the diluted insecticide or control for 15 s, and allowed to air-dry at room temperature. The leaves were placed individually inside sterile dishes. The third-day second-instar larvae of *G. daurica* were selected and used for the bioassay. Each treatment was replicated three times with each replication containing 10 test bugs. The control larvae were treated with 0.05% Triton X-100 sterile water. All second-instar larvae were kept at 24 ± 2 °C, RH 40% ± 10%, and L 14 h: D 10 h, and mortality was recorded after 48 h exposure. The individuals were considered dead if they failed to move or twitch slightly when touched with a brush.

5.4. Data Analysis

Mortality data were corrected using Abbott’s formula (1987) [58] and analyzed by Probit analysis using POLO-Plus version 2.0 [59]. The relative toxic index is the ratio of LC\textsubscript{50} values among different field populations.
Author Contributions: Conceptualization, W.Z. and Y.T.; methodology, W.Z., S.Y. and Y.T.; software, W.Z. and H.R.; validation, W.Z., H.R., F.S. and T.S.; formal analysis, W.Z. and Y.T.; investigation, W.Z., H.R., F.S. and T.S.; resources, S.Y. and Y.T.; writing—original draft preparation, W.Z. and Y.T.; writing—review and editing, W.Z., X.G. and Y.T.; supervision, Y.T. All authors have read and agreed to the published version of the manuscript.

Funding: This work was supported by the Inner Mongolian Science and Technology Key Project-2021 grant funded by the Inner Mongolian Chamber of Science and Technology (2021GG0108), Higher Education Young Scientific and Technology Talent Plan Fund (NJYT-19-B22), Research Start-up Funds for High-level Researchers in Inner Mongolia Agricultural University (NDYB2019-02) and Natural Science Foundation of Inner Mongolia (2020BS03033).

Institutional Review Board Statement: Not applicable.

Informed Consent Statement: Not applicable.

Data Availability Statement: Data supporting reported results are available upon request.

Acknowledgments: We thank Xianghuang Qi grassland workstation, Bayannur crop protection station, and the Scientific Observing and Experimental station for pests in Xilingol Rangeland, for allowing us to collect field populations of O. asiaticus, M. palpalis and G. daurica, and for their technical support during the research. We would like to thank Caroline Awmack for some professional editing.

Conflicts of Interest: The authors declare no conflict of interest.

References
1. Liu, X.; Wang, X.; Bai, M.; Shaw, J.J. Decrease in carabid beetles in grasslands of northwestern China: Further evidence of insect biodiversity loss. *Insects* 2021, 13, 35. [CrossRef] [PubMed]
2. Lu, Y.; Fu, B.; Wei, W.; Yu, X.; Sun, R. Major ecosystems in China: Dynamics and challenges for sustainable management. *Environ. Manag.* 2011, 48, 13–27. [CrossRef] [PubMed]
3. Zhou, W.; Li, J.; Yue, T. Grassland degradation restoration and constructing green ecological protective screen. In *Remote Sensing Monitoring and Evaluation of Degraded Grassland in China*; Zhou, W., Li, J., Yue, T., Eds.; Springer Geography: Singapore, 2020; pp. 125–138.
4. Zuo, X.; Zhao, H.; Zhao, X.; Guo, Y.; Yun, J.; Wang, S.; Miyasaka, T. Vegetation pattern variation, soil degradation and their relationship along a grassland desertification gradient in Horqin Sandy Land, northern China. *Environ. Geol.* 2009, 58, 1227–1237. [CrossRef]
5. Han, D.; Wang, G.; Xue, B.; Liu, T.; Yinglan, A.; Xu, X. Evaluation of semiarid grassland degradation in North China from multiple perspectives. *Ecol. Eng.* 2018, 112, 41–50. [CrossRef]
6. Wenhua, L. Degradation and restoration of forest ecosystems in China. *For. Ecol. Manag.* 2004, 201, 33–41. [CrossRef]
7. Kang, L.; Han, X.; Zhang, Z.; Sun, O.J. Grassland ecosystems in China: Review of current knowledge and research advancement. *Philos. Trans. R. Soc. B* 2007, 362, 997–1008. [CrossRef]
8. Peters, D.P. Plant species dominance at a grassland-shrubland ecotone: An individual-based gap dynamics model of herbaceous and woody species. *Ecol. Model.* 2002, 152, 5–32. [CrossRef]
9. Wang, Y.; Li, S.; Du, G.; Hu, G.; Zhang, Y.; Tu, X.; Zhang, Z. An analysis of the possible migration routes of *Oedaleus decorus asiaticus* Bey-Bienko (Orthoptera: Acrididae) from Mongolia to China. *Insects* 2022, 13, 72. [CrossRef]
10. Wang, Y.P.; Tu, X.B.; Lin, P.J.; Li, S.; Xu, C.M.; Wang, X.Q.; Reynolds, D.R.; Chapman, J.; Zhang, Z.H.; Hu, G. Migratory take-off behavior of the Mongolian grasshopper *Oedaleus asiaticus*. *Insects* 2020, 11, 416. [CrossRef]
11. Le Gall, M.; Overson, R.; Cease, A. A global review on locusts (*Orthoptera: Acrididae*) and their interactions with livestock grazing practices. *Front. Ecol. Evol.* 2019, 7, 263. [CrossRef]
12. Wang, D.; Ba, L. Ecology of meadow steppe in northeast China. *Rangel. J.* 2008, 30, 247–254. [CrossRef]
13. Tan, Y.; Zhou, X.R.; Pang, B.P. Reference gene selection and evaluation for expression analysis using qRT-PCR in *Galeruca daurica* (Joannis). *Bull. Entomol. Res.* 2017, 107, 359–368. [CrossRef] [PubMed]
14. Cui, B.; Huang, X.; Li, S.; Hao, K.; Chang, B.H.; Tu, X.; Pang, B.; Zhang, Z. Quercetin affects the growth and development of the grasshopper *Oedaleus asiaticus* (Orthoptera: Acrididae). *J. Econ. Entomol.* 2019, 112, 1175–1182. [CrossRef]
15. Zhang, L.; Lecoq, M.; Latchininsky, A.; Hunter, D. Locust and grasshopper management. *Annu. Rev. Entomol.* 2019, 64, 15–34. [CrossRef]
16. Dong, W.; Zhang, X.B.; Zhang, X.Y.; Wu, H.H.; Zhang, M.; Ma, E.B.; Zhang, J.Z. Susceptibility and potential biochemical mechanism of *Oedaleus asiaticus* to beta-cypermethrin and deltamethrin in the Inner Mongolia, China. *Pestic. Biochem. Physiol.* 2016, 132, 47–52. [CrossRef] [PubMed]
17. Li, H.; Zhang, Y.; Wang, G.; Lowry, A.; Huang, W.; Dong, Y.; Shang, S.; Luke, B. The effects of vegetation type on *Oedaleus decorus asiaticus* (Orthoptera: Acrididae) oviposition and hatching success. *Environ. Entomol.* 2021, 50, 790–794. [CrossRef] [PubMed]
18. Lomer, C.J.; Bateman, R.P.; Dent, D.; De Groote, H.; Douro–Kpindou, O.K.; Kooyma, C.; Ouambama, Z.; Peveling, R.; Kooyma, C.; Thomas, M. Development of strategies for the incorporation of biological pesticides into the integrated management of locusts and grasshoppers. *Agric. For. Entomol.* 1999, 1, 71–88. [CrossRef]

19. Liu, C.Z.; Yang, Y.B.; Ma, L.X. Spatial patterns of *Myrmeleotettix palpalis* and their application. *Pratac. Sci.* 2000, 17, 42–43.

20. Li, Z.W. Studies on the biological characteristics of *myrmeleotettix palpalis* on bayinbulake grasslands. *Plant Prot.* 2008, 34, 131–132.

21. Liu, C.Z.; Peng, G.H. Studies on the bionomics of *Myrmeleotettix palpalis*. *Acta Phytophytl.* Sin. 1999, 26, 153–156.

22. Cao, K.L.; Wang, Y.; Gao, Y.F.; Tan, S.Q.; Shi, W.P. Regulatory effects of vegetation on the behavior and population of grasshoppers. *J. Plant Prot.* 2021, 48, 59–64.

23. Kang, L.; Chen, Y.L. Trophic niche of steppe grasshop-pers. *Acta Entomol. Sin.* 1994, 37, 178–189.

24. Kang, L.; Chen, Y.L. Multidimensional analysis of resource utilization in assemblages of rangeland grasshoppers. *Entomol. Sin.* 1994, 1, 264–282.

25. Tan, Y.; Zhang, Y.; Huo, Z.J.; Zhou, X.R.; Pang, B.P. Molecular cloning of heat shock protein 10 (Hsp10) and 60 (Hsp60) cDNAs from *Galeruca daurica* (Coleoptera: Chrysomelidae) and their expression analysis. *Bull. Entomol. Res.* 2018, 108, 510–522. [CrossRef] [PubMed]

26. Duan, T.F.; Gao, S.J.; Wang, H.C.; Li, L.; Li, Y.Y.; Tan, Y.; Pang, B.P. MicroRNA let-7-5p targets the juvenile hormone primary response gene Krüppel homolog 1 and regulates reproductive diapause in *Galeruca daurica*. *Insect Biochem. Mol. Biol.* 2022, 142, 103727. [CrossRef]

27. Zhang, J.H.; Li, L.; Ni, L.; Li, Y.Y.; Pang, B.P. Expression profiling and functional analysis of candidate odorant receptors in *Galeruca daurica*. *Insects* 2022, 13, 563. [CrossRef] [PubMed]

28. Wang, H.C.; Li, L.; Li, Y.Y.; Tan, Y.; Pang, B.P. Expression profiling and functional analysis of candidate odorant receptors in *Galeruca daurica* (Coleoptera: Chrysomelidae) using qRT-PCR. *Entomol. Res.* 2021, 51, 393–402. [CrossRef]

29. Wojciechowska, M.; Stepnowski, P.; Gołębiewski, M. The use of insecticides to control insect pests. *ISJ—Invert. Surviv. J.* 2016, 13, 210–220.

30. Liu, J.W.; Zhang, Y.J.; Li, Y.J.; Wang, D.L.; Hou, F.J. Overview of grassland and its development in China. *Proc. Int. Grassl. Congr. Int. Rangel. Congr.* 2008, 6, 3–10.

31. Majak, J.I.; Lecoq, M.; Hunter, D.M. Preventive control and desert locust plagues. *Crop Prot.* 2008, 27, 1527–1533. [CrossRef]

32. Carruthers, R.I.; Onsager, J.A. Perspective on the use of exotic natural enemies for biological control of pest grasshoppers (*Orthoptera: Acrididae*). *Environ. Entomol.* 1993, 22, 885–903. [CrossRef]

33. Cease, A.J.; Elser, J.J.; Ford, C.F.; Hao, S.; Kang, L.; Harrison, J.F. Heavy livestock grazing promotes locust outbreaks by lowering plant nitrogen content. *Science* 2012, 335, 467–469. [CrossRef] [PubMed]

34. Mao, L.G.; Tu, X.B.; Liu, X.G.; Zhang, L.; Zhang, Y.; Zhu, L.Z.; Zhang, Y.Q.; Jiang, I.H. Advances and prospects of pesticides registered for controlling locusts and grasshoppers in China. *Mod. Agrochem.* 2021, 20, 1–7.

35. Cao, G.; Jia, M.; Zhao, X.; Wang, L.; Tu, X.; Wang, G.; Nong, X.; Zhang, Z. Effects of chlorantraniliprole on resistant *Oedaleus asiaticus* (*Deuteromycotina: Hyphomycetes*). *Bull. Entomol. Res.* 2021, 112, 671–684. [CrossRef] [PubMed]

36. Matthews, G.; Bateman, R.; Miller, P. *Pesticide Application Methods*; John Wiley & Sons: New York, NY, USA, 2014.

37. Guo, K.U.N.; Hao, S.G.; Sun, O.J.; Kang, L.E. Differential responses to warming and increased precipitation among three contrasting grasshopper species. *Glob. Change Biol.* 2009, 15, 2539–2548. [CrossRef]

38. Durst, D.G.; Donzelli, B.G.; Krasnoff, S.B.; Keyhani, N.O. Discovering the secondary metabolite potential encoded within entomopathogenic fungi. *Nat. Prod. Rep.* 2014, 31, 1287–1305. [CrossRef]
49. Guo, Y.; An, Z.; Shi, W. Control of grasshoppers by combined application of *Paranosema locustae* and an insect growth regulator (IGR) (cascade) in rangelands in China. *J. Econ. Entomol.* **2012**, *105*, 1915–1920. [CrossRef]

50. Chowański, S.; Kudlewska, M.; Marciniak, P.; Rosiński, G. Synthetic insecticides—is there an alternative? *Pol. J. Environ. Stud.* **2014**, *23*, 291–302.

51. Sparks, T.C.; Crouse, G.D.; Durst, G. Natural products as insecticides: The biology, biochemistry and quantitative structure-activity relationships of spinosyns and spinosoids. *Pest Manag. Sci.* **2001**, *57*, 896–905. [CrossRef]

52. Michaud, J.P.; Grant, A.K. IPM-compatibility of foliar insecticides for citrus: Indices derived from toxicity to beneficial insects from four orders. *J. Insect Sci.* **2003**, *3*, 18. [CrossRef]

53. Sparks, T.C.; Crouse, G.D.; Durst, G. Natural products as insecticides: The biology, biochemistry and quantitative structure-activity relationships of spinosyns and spinosoids. *Pest Manag. Sci.* **2001**, *57*, 896–905. [CrossRef]

54. Chowański, S.; Kudlewska, M.; Marciniak, P.; Rosiński, G. Synthetic insecticides—is there an alternative? *Pol. J. Environ. Stud.* **2014**, *23*, 291–302.

55. Sparks, T.C.; Crouse, G.D.; Durst, G. Natural products as insecticides: The biology, biochemistry and quantitative structure-activity relationships of spinosyns and spinosoids. *Pest Manag. Sci.* **2001**, *57*, 896–905. [CrossRef]

56. Michaud, J.P.; Grant, A.K. IPM-compatibility of foliar insecticides for citrus: Indices derived from toxicity to beneficial insects from four orders. *J. Insect Sci.* **2003**, *3*, 18. [CrossRef]

57. Sparks, T.C.; Crouse, G.D.; Durst, G. Natural products as insecticides: The biology, biochemistry and quantitative structure-activity relationships of spinosyns and spinosoids. *Pest Manag. Sci.* **2001**, *57*, 896–905. [CrossRef]

58. Abbott, W.S.A. A method of computing the effectiveness of an insecticide. *J. Econ. Entomol.* **1925**, *18*, 265–267. [CrossRef]

59. Robertson, J.L.; Savin, N.E.; Preisler, H.K.; Russell, R.M. *Bioassays with Arthropods*, 2nd ed.; CRC Press: Boca Raton, FL, USA, 2007; ISBN 9780849323317.