Effectiveness of Biochar Obtained from Corncob for Immobilization of Lead in Contaminated Soil

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Introduction

Contamination of soils by both organic and inorganic pollutants is an issue worldwide, and environmentally friendly alternatives for addressing this problem are being investigated.1 A number of new alternative techniques utilize an in-situ, low-invasive approach involving plants (with or without chemical additives) to reduce contaminant transfer to the environment by direct extraction of pollutants (clean up) or by soil stabilization (using biological or chemical processes). Collectively, these techniques are sometimes referred to as “gentle” remediation options.2

Biochar is a solid material that is rich in carbon and is synthesized by hydrothermal carbonization or by slow pyrolysis of biomass.3 These processes produce biochar with a remarkable alkaline nature that is favorable for the treatment of acidic soils. The processes principally involve the thermal decomposition of biomass such as oil palm, cottonseed husk, orange peel, bamboo, and various organic wastes under anaerobic conditions.4-8 The application of biochar as a means of remediation and soil strengthening has been studied over the last decade due to its efficiency and cost effectiveness. These days, over one thousand studies around biochar and soil enhancement are published each year, and scientific interest is growing.9-12 Biochar’s physicochemical properties, such as porosity, surface area, pH, conductivity, and structure are determining factors in its impact and interaction with soil that help to increase crop yields and carbon sequestration, reduce soil greenhouse gases, and favor the immobilization of organic and inorganic (including metallic) pollutants.9,11,13-21

Background. Recent studies have explored the potential for using biochar as a soil amendment in agriculture. However, it can also be used as a gentle remediation option for contaminant reduction. Biochar is a by-product obtained from the pyrolysis of biomass (organic matter). It is known for its long-lasting chemical properties, wide surface area values, and carbon-richness, which make it an efficient method for the immobilization of organic and inorganic contaminants such as heavy metals.

Objective. The aim of the present study was to analyze the efficiency of biochar, obtained from the gasification of corncob, for the immobilization of lead in contaminated soils.

Methods. In the present study, biochar from corncob was used as an amendment for soil contaminated with lead (extracted from the municipality of Malambo, Colombia) in order to estimate its ability to immobilize leaching lead. A comparison laboratory test applied a modified biochar produced with a 10% hydrogen peroxide chemical treatment. In addition, a pot experiment was done with both biochar by sowing seeds of Pennisetum clandestinum for 33 days. During this period, plant growth was measured for the different amendments of biochar concentrations.

Results. Laboratory tests indicated that unmodified biochar obtained a maximum retention of 61.46% of lead, while the modified biochar obtained only 44.53% retention. In the pot experiments, the modified biochar indicated high germination and growth of seeds (up to 89.8%).

Conclusions. Although the lead immobilization in soil was positive for both cases, the use of soil with high concentrations of lead (167.62 g/kg) does not indicate biochar’s effectiveness for purposes of comparison with the current United States Environmental Protection Agency (USEPA) limit value (400 ppm for bare soil in urban play areas). Therefore, further studies are recommended using soil with lower lead concentration levels.

Competing Interests. The authors declare no competing financial interests. One author is an employee of Pure Earth.

Keywords. biochar, soil remediation, gasification, contamination, lead, pH, moisture, retention, adsorption, carbon sequestration, heavy metals
This research aims to analyze the efficiency of biochar obtained from the gasification of corncob (after the corn kernels have been removed) with the immobilization of lead in contaminated soil.

**Methods**

The tests were performed on soil contaminated with high levels of lead from the municipality of Malambo, located on the north coast of Colombia. Additionally, a chemical modification of biochar was performed, and the results obtained from both types of biochar were compared. Physicochemical tests were carried out in order to evaluate the changes generated in the soil using the two types of biochar. In addition, a *Pennisetum clandestinum* pot experiment was conducted parallel to the previously mentioned tests to analyze the effect of biochar on plant growth.

Hydrogen peroxide was used to produce the modified biochar sample, which increased the functional groups containing oxygen and aided metal sorption. In comparison with other techniques, the hydrogen peroxide modification is cost effective, easily accessible, the decomposition products H₂O and O₂ are environmentally friendly, and at a 10% concentration it has a greater absorption compared to commercial alternatives. Chemical activation methodologies with potassium hydroxide, carbon dioxide and steam current physical techniques require high temperatures, approximately 800°C, which increases the process costs and risks.

**Sampling**

Thirty (30) kg of soil was collected superficially from an abandoned lead smelter in La Bonga Village in the Malambo municipality. The location is a public health concern due to the associated lead poisoning cases in the surrounding community. The biochar residue was obtained from the gasification of corncob, which was used to produce renewable energy.

**Soil characterization**

Soil texture and structure were determined using a physical analysis procedure provided by the Ministry of Agriculture and Forestry of Alberta, Canada. A sample of 100 g of homogenized soil was required for the analysis. Subsequently, a chemical analysis provided an adequate characterization of the soil, as well as values of electrical conductivity, pH, volatile solids, and humidity. These properties were measured according to Banos *et al.* Moisture was measured using the gravimetric method.

The concentration of lead in the soil indicates the amount of lead available for each planting modification. Lead concentration was determined through an analysis of total metals by inductively coupled plasma mass spectrometry, which employs a microwave-assisted acid digestion method according to standards proposed by the United States Environmental Protection Agency (USEPA).

**Biochar characterization**

The conditions to conduct the pyrolysis process include raising the temperature range to 130-600°C and forming part of the gasification process of the corncob. The pyrolysis process consisted of four principal stages: first, drying the biomass inside the hopper of the gasifier; second, using tar removal to eliminate the mass percentage that was not considered in the design; third, provide devolatilization or decomposition of the biomass in its constituent elements, and fourth, biomass gasification. Before gasification, the sample was brought to a humidity of less than 30% and a particle size between 1 and 4 cm. The equipment used was a fixed-bed, downdraft gasifier (ALL Power Labs, California, USA). The pyrolysis process is detailed in a thesis on renewable energy from the University of La Sabana by Martinez. The white rachis of corn used in the present study is mainly composed of cellulose (40-50%), hemicellulose (20-30%), and lignin (10-40%).

After gasification, the biochar was sieved in order to avoid large granules or chunks. The final result was a homogeneous fine powder.

**Biochar chemical and physical analysis**

The physical-chemical biocarbon analysis used 10 g of biochar and 200 ml of water. Previously suggested methods were used to determine the electrical conductivity and pH of the biochar. The surface area was calculated by the Brunauer-Emmett-Teller method, which deducts the surface area by desorption and

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**Abbreviations**

| Abbreviation | Description |
|--------------|-------------|
| USEPA        | United States Environmental Protection Agency |
adsorption of N2 at 77 K. \(^{30}\) Lastly, the final analysis was carried out following ASTM D5373-14, method A. \(^{31,32}\)

**Biochar modification**

Biochar modification was performed by dissolving 261 g of biochar in 1740 ml of hydrogen peroxide at 10%, giving rise to an exothermic reaction. The mixture was then left to stand for 2 hours at a temperature of 22°C. Finally, it was washed with distilled water and dried in an oven at 80°C. \(^{19}\)

**Execution of the sowing test (pot experiment)**

*Pennisetum clandestinum*, a grass-forming species that can spread progressively, was chosen for sowing in the pot experiment. *Pennisetum clandestinum* adapts easily to humid tropics or subtropics, especially at higher elevations and in high fertility soils. \(^{33}\)

Table 1 presents the baseline soil characterization results, unmodified biochar treatment results, and the modified biochar treatment results for non-modified biochar, both with equal measures. Each were filled with 80 g of contaminated soil and a specific percentage of biochar, homogenized by sieving. The biochar-soil mixtures consisted of 0%, 1%, 1.5%, 2.5%, 4.5% and 7% biochar concentrations and in triplicates, as shown in Figure 1.

Application began by depositing 3/8\(^{th}\) of the mixture’s total volume. Then, the sowing was carried out by the furrow method, which involved placing the *Pennisetum clandestinum* seeds in a linear pattern, followed by covering the seeds with the remaining mixture. \(^{34}\)

Spatial uniformity factors and depth were taken into account for sowing. \(^{35-37}\)

Twenty (20) seeds of *Pennisetum clandestinum* were deposited and evenly spaced. The application was then repeated, starting with depositing 3/8\(^{th}\) of the mixture’s total volume and followed by 20 more seeds spread on the surface before being covered with the remaining mixture. Therefore, each pot had a total of 40 seeds. Each pot was watered every 24 hours and provided 50 ml of water per cell. The pots were placed under a controlled laboratory environment. Follow-up for the two sowing tests occurred over 33 days. The sowing parameters were the same for both biochar types.

**Detection of lead reduction**

The concentration of lead reduction after biochar amendment was determined through an analysis of total metals by inductively coupled plasma mass spectrometry, which employs a microwave assisted acid digestion method according to standards proposed by the USEPA. \(^{27}\)

**Results**

Table 1 presents the baseline soil characterization results, unmodified biochar treatment results, and the modified biochar treatment results for non-modified biochar, both with equal measures. Each were filled with 80 g of contaminated soil and a specific percentage of biochar, homogenized by sieving. The biochar-soil mixtures consisted of 0%, 1%, 1.5%, 2.5%, 4.5% and 7% biochar concentrations and in triplicates, as shown in Figure 1.

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Two pots were chosen for sowing experiments and growth monitoring, one pot for modified biochar and other for non-modified biochar, both with equal measures. Each were filled with 80 g of contaminated soil and a specific percentage of biochar, homogenized by sieving. The biochar-soil mixtures consisted of 0%, 1%, 1.5%, 2.5%, 4.5% and 7% biochar concentrations and in triplicates, as shown in Figure 1.
(modified with 10% hydrogen peroxide).

The non-modified biochar contained carbon, hydrogen, nitrogen and oxygen in percentages of 76.3%, 2.18%, 10.53%, and 1.31%, respectively. The non-modified biochar was obtained from corncob organic matter under a pyrolysis temperature between 130°C and 600°C.

Lead availability in soil as a function of biochar concentration

The first pot experiment tested unmodified biochar and the second experiment tested modified biochar. The results of both are presented in Figure 2.

Statistical analysis indicated differences between biochar types. The unmodified biochar treatment presented a lower soil lead concentration. However, the concentration of lead from the modified biochar was still acceptable as the variation coefficient for each percentage was lower than 25%, as shown in Table 2.

Sowing results

The 33-day growth monitoring showed behavior as an exponential function of time for each biochar concentration. Regarding the unmodified biochar pot experiment, it was observed that as the biochar concentration increased, there was a significant negative growth effect (Figure 3). Conversely, increases in modified biochar concentrations showed a positive effect on growth with the 7% concentration providing the best results. This result is in contrast to the unmodified biochar results where the 7% concentration of unmodified biochar provided the lowest growth.

Seed germination occurred after 12 days in the unmodified biochar concentrations and after 10 days in the modified biochar concentrations.

Development significantly improved in cells with 1.5%, 2.5%, 4.5%, and 7% concentrations of modified biochar, and growth was improved by 4.6%, 34.2%, 52.1%, and 89.8%, respectively. The cell containing 1% of modified biochar showed low growth with just a 4.7% difference in height.

Regarding lead contamination, sowing with unmodified biochar caused a reduction in lead concentration of up to 61.46%. The reduction in lead concentration was greater in the modified biochar concentrations.
contamination using modified biochar was 44.53%. Although the reduction percentages are significant, it was not possible to bring lead contamination levels below the USEPA permissible standards of "400 parts per million (ppm) of lead in bare soil in children’s play areas or 1200 ppm average for bare soil in the rest of the yard."\textsuperscript{38,39} These standards were determined in 2001 according to the maximum level of lead to which a child could be safely exposed, given that children represent the most vulnerable population in terms of health.\textsuperscript{39,40}

**Discussion**

The unmodified biocarbon had a higher metal retention of 61.46% compared to the modified biocarbon.
retention of 44.53%. However, the desired outcomes of the pot experiments were inversely related: the modified biochar provided higher plant growth while the unmodified biochar provided lower plant growth. Yet both types of biocarbon generated positive results in the soil’s physical and chemical properties and in sowing growth. Generally, the type of biocarbon dictates the resulting inhibition of microbial activity that causes the immobilization of nitrogen that is vital for plant growth. It is important to note that the same raw material can produce different types of biochar, each with a different nitrogen content that impacts the biochar-soil interaction.

The biochar-treated soil was composed of single grain, structureless soil. Characteristics of structureless soil include increased runoff and high risk of water erosion. Use of this soil could be the cause for the limited growth seen in the present study. Biochar has a granular structure and can provide a better growing environment for plant roots, likely due to its high carbon content that contributes to the plant life cycle. Carbon provides greater water retention and permeability in soil, so the addition of biochar can substantially improve water retention capacity.

Modified biochar increased pH and decreased the solubility of metals (with the exception of metalloids). Hence, biochar with basic pH levels is known for yielding higher crop productivity in acidic soils. Soil alkalization is vital to understanding the growth and immobilization of pollutants, because the pH of biochar directly affects pH levels in soil. For a highly alkaline biochar and acidic soil sample, a limestone amendment effect normally occurs, as the acidity of the soil will decrease, leading to a significant increase in crop growth. However, not all types of biochar generate the same growth effect because the results are related to the type of species being cultivated.

The best growing conditions were observed in cases of moderate or low biochar addition, similar to the results of previous studies. The conditions (pH, conductivity, immobilization capacity, surface area) of growth are improved by modification, as observed in this study’s modified biochar treatment results (Figure 4 and 5). Different remediation patterns were observed with unmodified biochar where pH increased, but growth decreased considerably. This may have been due to biochar’s high adsorption capacity, capturing essential nutrients from the plant, such as phosphorus and sulfur. Therefore, the availability of these nutrients in the soil may decrease and adversely affect plant development.

Furthermore, growth was inhibited by the unmodified biochar amended-soil’s high levels of conductivity, where conductivity deviated between acceptable growth values of 0-0.8 ms/cm.

Table 3 indicates that the biochar modification process contributed to decreased and regulated soil conductivity compared to the unmodified biochar, where conductivity values were too high. Consequently, the latter substrate was not a propitious environment for growth. Organic matter content is an important variable in the development and absorption of nutrients, where the appropriate level for sowing grasses is between 8% and 12%, and the minimum soil organic matter content for basic plant development is 2%. However, this is not ideal for sowing pastures. The soil organic matter content for the current study’s soil sample was 3.9%, and although this is not the

| Biochar (%) | pH    | **pH | Electrical conductivity (mS/cm) | **Electrical conductivity (mS/cm) | Lead (g/kg) | **Lead (g/kg) | Lead retention (%) | **Lead retention (%) | Maximum growth (mm) | **Maximum growth (mm) |
|------------|-------|------|-------------------------------|-------------------------------|------------|--------------|------------------|--------------------|---------------------|---------------------|
| 0.00       | 5.53  | 5.53 | 0.73                          | 0.73                          | 167.62     | 167.62       | -                | -                  | 150                 | 126                 |
| 1.00       | 7.36  | 7.89 | 2.27                          | 1.15                          | 107.45     | 100.81       | 35.90            | 39.86              | 140                 | 142                 |
| 1.50       | 8.01  | 8.01 | 3.38                          | 1.54                          | 94.48      | 130.62       | 43.63            | 22.08              | 130                 | 140                 |
| 2.50       | 8.87  | 8.98 | 4.43                          | 1.96                          | 107.80     | 141.75       | 35.69            | 15.43              | 80                  | 147                 |
| 4.50       | 9.40  | 9.93 | 5.15                          | 2.20                          | 113.77     | 110.44       | 32.13            | 34.11              | 70                  | 160                 |
| 7.00       | 9.72  | 10.14| 9.50                          | 3.23                          | 64.60      | 92.98        | 61.46            | 44.53              | 31                  | 160                 |

** The results obtained with modified biochar.
ideal value for grass growth, it meets the minimum percentage of organic matter for growth to occur.

Other studies of biochar chemical modification show the potential for increased biochar surface area, which is an important indicator of its adsorption capacity. A study by Ahmad et al. showed that a wheat straw biochar sample had a surface area of 4.5 m²/g. A graphene modification increased the sample's surface area to 17.3 m²/g and, in turn, improved mercury retention by 31.6%. Thus, the surface area of the graphene-modified biochar was approximately 3.84 times greater than that of the unmodified biochar. This outcome was reflected in a hydrogen peroxide modification, where the surface area was approximately 2.45 times greater than that of the unmodified biochar.

The relationship between oxygen and carbon indicates how aromatic or hydrophilic the surface of biochar can be. The ratio obtained in the present study was 0.138. Thus, the surface of the biochar is more aromatic than hydrophilic, which is due to a greater carbon extension and loss of functional groups that present a polar nature at high temperatures. Evidence of this was provided by the elemental analysis that indicated the high aromatic carbon content. Both the organic matter of origin and the pyrolysis conditions are significant in the final carbon concentration.

Biochar used in this study had a carbon concentration of 76.3%, which is higher than other types of biochar. For example, the biochar of spruce wood has a 51.21% carbon concentration and biochar from corn waste at a temperature of 350°C contains 67.5% carbon, however, at 600°C it has a higher content (79.0%).

In a study by Ahmad et al. a comparison was done between several biochar feedstock from broiler waste, buffalo weed, canola waste, cottonseed coatings, orange peel shells, peanut shells, poultry manure, sewage sludge, and wood waste, among others. Elemental analysis values were in the range of 20.19% - 95.30% for carbon, 0.42%-7.25 for hydrogen, 0.01%-46.80% for oxygen and 0.04%-10.21% for nitrogen. Thus, the obtained biochar of corn cob values are within the expected ranges.

The results of soil lead concentration were favorable with regard to lead retention capacity when using unmodified biochar (Table 3). The retention of lead from modified biochar did not present any significant changes to justify chemical modification costs, as the samples were not below the permissible limit (400 ppm). This was due to the extremely high contamination levels that were found in the study’s soil.

Soil conductivity dropped by 50.66%, from 2.27 mS/cm to 1.15 mS/cm (in cells with soil and biochar at 1%). Both biochars created healthy soil conditions (0-0.8 mS/cm).

Conclusions

Although the use of biochar as a soil amendment is still considered an option for strengthening the organic matter and increasing the growth of species, there is not enough scientific evidence to support its use as a remediation method in contaminated soils or as an alternative intervention.

Further studies are needed using soils with more environmentally viable concentrations of lead and biochar comprised of different types and combinations of biomass.

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References

1. Mench M, Lepp N, Bert V, Schultze-Guembel JP, Gawronski SW, Schoder P, Vangronsveld J. Successes and limitations of phytotechnologies at field scale: outcomes, assessment and outlook from COST action 859. J Soils Sediment [Internet]. 2010 [cited 2019 Jun 19];10(6):1039-70. Available from: https://doi.org/10.1007/s11368-010-0190-x. Subscription required to view.
2. Onwubuya K, Candy A, Puschreiter M, Kumpiene J, Bone B, Greaves J, Teasdale P, Mench M, Thusto P, Mikhailovsky S, Waite S, Fries-Hanli W, Marschner B, Muller I. Developing decision support tools for the selection of “gentle” remediation approaches. Sci Total Environ [Internet]. 2009 Dec 1 [cited 2019 Jun 19];407(24):6132-42. Available from: https://doi.org/10.1016/j.scitotenv.2009.08.017. Subscription required to view.
3. Tan X, Liu Y, Zeng G, Wang X, Hu X, Gu Y, Yang Z. Application of biochar for the removal of pollutants from aqueous solutions. Chemosphere [Internet]. 2015 Apr [cited 2018 Jan];125:70-85. Available from: https://doi.org/10.1016/j.chemosphere.2014.12.058. Subscription required to view.
4. Ahmad M, Rajapaksha AU, Lim JE, Zhang M, Bolan N, Mohan D, Vithanage M, Lee SS, Ok YS. Biochar as a sorbent for contaminant management in soil and water: a review. Chemosphere [Internet].
Biochar from Corncob for Immobilization of Lead in Contaminated Soil

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13. Brewer CE, Chuan VJ, Masidello CA, Gonsenmerm H, Gao X, Dugan K, Driver LE, Panizcchi P, Zygonarakis K, Davies CA. New approaches to measuring biochar density and porosity. BioMass Bioenergy [Internet]. 2014 Jul [cited 2018 Jan];66:176-85. Available from: https://doi.org/10.1016/j.biombioe.2014.03.059 Subscription required to view.

14. Rebolledo AE, Lopez GP, Moreno CH, Collado JL, Alves JC, Pacheco EV, Barra JD. Biocarbon (biochar) I: Naturaleza, historia, fabricación y uso en el suelo. [Biochar (biochar) I: Nature, history, manufacture and use in soil]. Terra LatinoAmericana [Internet]. 2016 [cited 2018 Feb];34(2):367-82. Available from: http://www.redalyc.org/articulo. oxid@57346617009. Spanish.

15. Wu D, Senbayram M, Zhang H, Ugarlar F, Aydemir S, Bruggemann N, Kunyakov Y, Bol R, Blagodatskaya E. Effect of biochar origin and soil pH on greenhouse gas emissions from sandy and clay soils. Appl Soil Ecol [Internet]. 2018 Aug [cited 2019 Jun 19];129:121-7. Available from: https://doi.org/10.1016/j.apsoil.2018.05.009 Subscription required to view.

16. Xiao R, Wang JJ, Gaston LA, Zhou B, Park JH, Li R, Dodla SK, Zhang Z. Biochar produced from mineral salt-imregnated chicken manure: fertility properties and potential for carbon sequestration. Waste Manag [Internet]. 2018 Aug [cited 2019 Jun 19];78:802-10. Available from: https://doi.org/10.1016/j.wasman.2018.06.047 Subscription required to view.

17. Li H, Dong X, da Silva EB, de Oliveira LM, Chen Y, Ma LQ. Mechanisms of metal sorption by biochars: biochar characteristics and modifications. Chemosphere [Internet]. 2017 Jul [cited 2018 Feb];178:466-78. Available from: https://doi.org/10.1016/j.chemosphere.2017.03.077 Subscription required to view.

18. Rajapaksha AU, Chen SS, Tsang DC, Zhang M, Vithanahe M, Mandal S, Gao B, Bolan NS, Ok YS. Engineered/designer Biochar for contaminant removal/immobilization from soil and water: potential and implication of biochar modification. Chemosphere [Internet]. 2016 Apr [cited 2017 Nov];148:276-91. Available from: https://doi.org/10.1016/j.chemosphere.2016.01.043 Subscription required to view.

19. Xue Y, Gao B, Yao Y, Inyang M, Zhang M, Zimmerman AR, Ro KS. Hydrogen peroxide modification enhances the ability of biochar (hydrochar) produced from hydrothermal carbonization of peanut hull to remove aqueous heavy metals: batch and column tests. Chem Eng J [Internet]. 2012 Aug 15 [cited 2018 Jan];200-202:673-80. Available from: https://doi.org/10.1016/j.cej.2012.06.116 Subscription required to view.

20. Zhou Y, Gao B, Zimmerman AR, Fang J, Sun Y, Cao X. Sorption of heavy metals on chitosan-modified biochars and its biological effects. Chem Eng J [Internet]. 2013 Sep [cited 2018 Jan];231:512-8. Available from: https://doi.org/10.1016/j.cej.2013.07.036 Subscription required to view.

21. Hiller E, Fargasa A, Zemenova L, Bartal M. Influence of wheat ash on the MCPA immobilization in agricultural soils. Bull Environ Contam Toxicol [Internet]. 2008 Sep [cited 2018 Mar];81(3):285-8. Available from: https://doi.org/10.1007/s00128-008-9400-2 Subscription required to view.

22. Wang Y, Liu R. H2O2 Treatment enhanced the heavy metals removal by manure biochar in aqueous solutions. Sci Total Environ [Internet]. 2018 Jul 1 [cited 2019 Jun 19];628-629:1139-1148. Available from: https://doi.org/10.1016/j.scitotenv.2018.02.137 Subscription required to view.

23. Zuo X, Liu Z, Chen M. Effect of H2O2 concentrations on copper removal using the modified hydrothermal biochar. Bioresour Technol [Internet]. 2016 May [cited 2019 Jun 19];207:262-7. Available from: https://doi.org/10.1016/j.biortech.2016.02.032 Subscription required to view.

24. Field assessment [Internet]. In: Nutrient management: planning guide. Alberta, Canada: Ministry of Agriculture and Forestry; [cited 2018 Jul 16]. Chapter 3.1. Available from: https://www1.agric.gov.ab.ca/$department/deptdocs.nsf/all/epw1920/$FILE/3-1.pdf

25. Banos SM, Galvan CP, Baraja CG. Prácticas de laboratorio de remediación de suelos y acuíferos [Laboratory practices for soil and aquifer remediation]. Mexico City, Mexico: Instituto Politécnico Nacional; 2011 Aug. 189 p. Spanish.
Characterization of products from hydrothermal treatments of cellulose. Energy [Internet]. 2012 Jun [cited 2017 Nov;42(1):457-65. Available from: https://doi.org/10.1016/j.energy.2012.03.023 Subscription required to view.

29. Dhyni V, Bhaskar T. A comprehensive review on the pyrolysis of lignocellulosic biomass. Renew Energy [Internet]. 2018 Dec [cited 2018 Jan];129(part B):695-716. Available from: https://doi.org/10.1016/j.renene.2017.04.035 Subscription required to view.

30. Girado L, García V, Moreno JC. Caracterización superficial en fase g y líquida de carbonos activados [Superficial characterization in gas and liquid phase of activated carbon]. Revista de Ingeniería. 2008 May;27:7-16. Spanish.

31. Reporte de Análisis. Bogota, Colombia: SGS Colombia S.A.S.; 2017. Report No.: 1704298669AR.

32. ASTM. Standard Test Methods for Determination of Carbon, Hydrogen and Nitrogen in Analysis Samples of Coal and Carbon in Analysis Samples of Coal and Coke [Internet]. Revista de Ingeniería. 2014 [cited 2018 Jan]. Available from: www.astm.org.

33. Pennisetum clandestinum: hochst. ex Chiov., Poaceae [Internet]. Hawaii: Pacific Island Ecosystems at Risk; 1999 Jan 1 [updated 2011 Aug 22; cited 2019 Jan 21]. [about 5 screens]. Available from: http://www.hear.org/pier/species/pennisetum_clandestinum.htm

34. Franco LH, Calero DQ, Duran CV. Manual de establecimiento de pasturas [Internet]. Bogota, Colombia: Universidad Nacional de Colombia; 2007 [cited 2017 Nov;32 p. Available from: http://bibdigital.unal.edu.co/5053/1/9789584411761.pdf Spanish.

35. Haegle M, Arjharn W. The effects of cultivation methods and planting season on biomass yield of Napier grass (Pennisetum purpureum Schumach.) under rainfed conditions in the northeast region of Thailand. Field Crop Res [Internet]. 2017 Dec [cited 2018 Mar];214:359-64. Available from: https://doi.org/10.1016/j.fcr.2017.09.027 Subscription required to view.

36. Leverkus AB, Rojo M, Castro J. Habitat complexity and individual acorn protectors enhance the post-fire restoration of oak forests via seed sowing. Ecol Eng [Internet]. 2015 Oct [cited 2017 Nov;83:276-80. Available from: https://doi.org/10.1016/j.ecoleng.2015.06.033 Subscription required to view.

37. Gross KL. A comparison of methods for estimating seed numbers in the soil. J Ecol [Internet]. 1990 Dec [cited 2017 Nov;78(4):1079-93. Available from: https://www.jstor.org/stable/2260953 Subscription required to view.

38. EPA announces tough new standards for lead [Internet]. Washington D.C.: US Environmental Protection Agency; 2010 Dec 26 [cited 2018 Feb]. [about 2 screens]. Available from: URL: https://archive.epa.gov/epanews/newsreleases/150103b59f35fcb5852566f1005feb9.html

39. Lead toxicity [Internet]. Atlanta, GA: Agency for Toxic Substances and Disease; 2017 Jun 12 [cited 2018 March]. 185 p. Available from: https://www.atsdr.cdc.gov/csr/lead/docs/CSEM-Lead_toxicity_508.pdf

40. Environmental Protection Agency (US). Lead; identification of dangerous levels of lead. Final rule. Fed Regist. 2001 Jan 5;66(4):1206-40.

41. Glaser B, Lehmann J, Steiner C, Nehr T, Yosuf M, Zech W. Potential of pyrolyzed organic matter in soil amelioration. In: Juren J, editor. Sustainable utilization of global soil and water resources. 12th ISCO Conference; 2002 May 26-31; Beijing, China. Beijing, China: Ministry of Water Resources; 2002. p. 421-7.

42. Kolb SE, Forganich KJ, Dornbush ME. Effect of charcoal quantity on microbial biomass and activity in temperate soils. Soil Sci Soc Am J. 2009 Jul-Aug;73(4):1173-81.

43. Deenik J, McClean T, Uchara G, Antal MJ, Campbell S. Charcoal volatile matter content influences plant growth and soil nitrogen transformations. Soil Sci Soc Am J. 2010 Jul;74(4):1259-70.

44. De Grys E, Cullen M, Durschinger L. Evaluation of the opportunities for generating charcoal offsets from soil sequestration of biochar [Internet]. San Francisco, CA: Terra Global Capital; 2010 Apr [cited 2018 Feb]. 99 p. Available from: https://www.terraglobalcapital.com/wp-content/uploads/2009/03/Soil_Sequestration_Biochar_Issue_Paper1.pdf

45. Sheng Y, Zhu L. Biochar alters microbial community and carbon sequestration potential across different soil pH. Sci Total Environ [Internet]. 2018 May 1 [cited 2018 Aug];622-623:1391-9. Available from: https://doi.org/10.1016/j.scitotenv.2017.11.337 Subscription required to view.

46. Beeley L, Moreno E, Pellet G, Carrijo L, Sizmur T. Biochar and heavy metals. In: Lehmann J, Joseph S, editors. Biochar for environmental management: science, technology and implementation. 2nd ed. Earthscan: London, UK; 2015. Chapter 20.

47. Gale NV, Sackett TE, Thomas SC. Thermal treatment and leaching of biochar alleviates plant growth inhibition from mobile organic compounds. PeerJ [Internet]. 2016 Aug 25 [cited 2019 Jun 19];eArticle e2385 [19 p.]. Available from: https://doi.org/10.7717/peerj.2385

48. Gale NV, Thomas SC. Dose–dependence of growth and ecophysiological responses of plants to biochar. Sci Total Environ [Internet]. 2019 Mar 25 [cited 2019 Jun 19];658:1344-54. Available from: https://doi.org/10.1016/j.scitotenv.2018.12.239 Subscription required to view.

49. Liu X, Zhang A, Ji C, Joseph S, Biao R, Li L, Pan G, Paz-Ferreiro J. Biochar’s effect on crop productivity and the dependence on experimental conditions-a meta-analysis of literature data. Plant Soil [Internet]. 2013 Dec [cited 2019 Jun 19];373(1-2):583-94. Available from: https://doi.org/10.1007/s11104-013-1806-x Subscription required to view.

50. Lehmann J, Rilling MC, Thies J, Masiello CA, Hockaday WC, Crowley D. Biochar effects on soil biota - a review. Soil Biol Biochem [Internet]. 2011 Sep [cited 2017 Dec];43(9):1812-36. Available from: https://doi.org/10.1016/j.soilbio.2011.04.022 Subscription required to view.

51. Schulz H, Glaser B. Effects of biochar compared to organic and inorganic fertilizers on soil quality and plant growth in a greenhouse experiment. J Plant Nutr Soil Sci [Internet]. 2012 Jun [cited 2018 Jan];175(3):410-22. Available from: https://doi.org/10.1002/jpln.201100143 Subscription required to view.

52. Zhang G, Guo X, Zhu Y, Han Z, He Q, Zhang F. Effect of biochar on the presence of nutrients and ryegrass growth in the soil from an abandoned indigenous coking site: the potential role of biochar in the revegetation of contaminated site. Sci Total Environ [Internet]. 2017 Dec 1 [cited 2018 March];601-602:469-477. Available from: https://doi.org/10.1016/j.scitotenv.2017.05.318 Subscription required to view.

53. Schulz H, Dunst G, Glaser B. Positive effects of composted biochar on plant growth and soil fertility. Agron Sustain Dev [Internet]. 2013 Oct [cited 2017 Nov];33(4):817-27. Available from: https://doi.org/10.1007/s11359-013-1580-0 Subscription required to view.

54. Beeley L, Moreno-Jimenez E, Gomez-Eyles JL, Harris E, Robinson B, Sizmur T. A review of biochars potential role in the remediation, revegetation and restoration of contaminated soils. Environ Pollut [Internet]. 2011 Dec [cited 2018 Feb];159(12):3269-82. Available from: https://doi.org/10.1016/j.envpol.2011.07.023 Subscription required to view.

55. Jones BE, Haynes RJ, Phillips IR. Effect of amendment of bauxite processing sand with organic
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Biochar from Corncob for Immobilization of Lead in Contaminated Soil material on its chemical, physical and microbial properties. J Environ Manag [Internet]. 2010 Nov [cited 2018 Jan];91(11):2281-8. Available from: https://doi.org/10.1016/j.jenvman.2010.06.013 Subscription required to view.

56. Uchimiya M, Lima IM, Klasson KT, Wartelle LH. Contaminant immobilization and nutrient release by biochar soil amendment: roles of natural organic matter. Chemosphere [Internet]. 2010 Aug [cited 2018 Jan];80(8):935-40. Available from: https://doi.org/10.1016/j.chemosphere.2010.05.020 Subscription required to view.

57. Yao Y, Gao B, Inyang M, Zimmerman AR, Cao X, Pullammanappallil P, Yang L. Biochar derived from anaerobically digested sugar beet tailings: characterization and phosphate removal potential. Bioreour Technol [Internet]. 2011 May [cited 2018 Jan];102(10):6273-8. Available from: https://doi.org/10.1016/j.biortech.2011.03.006 Subscription required to view.

58. Allen DE, Singh BP, Dalal RC. Soil health indicators under climate change: a review of current knowledge. In: Singh BP, Cowie AL, Chan KY, editors. Soil health climate change [Internet]. Berlin, Germany: Springer; 2011 [cited 2018 May]. p. 25-45 (Soil biology; vol. 29). Available from: https://doi.org/10.1007/978-3-642-20256-8_2 Subscription required to view.

59. Arnold SL, Doran JW, Schemper JS, Wienhold BJ, Ginting D. Portable probes to measure electrical conductivity and soil quality in the field. Commun Soil Sci Plant Anal [Internet]. 2005 [cited 2018 Jan];36(15-16):2271-87. Available from: https://doi.org/10.1080/0010362050186689 Subscription required to view.

60. Zambrano G, Apraez JE, Navia JF. Relación de las propiedades del suelo con variables bromatológicas de pastos, en un sistema lechero de nariño [Relationship between soil properties and bromatological variables of grasses in a dairy system of Nariño]. Revista Cienc Agrícolas [Internet]. 2014 Jul-Dec [cited 2018 Feb];31(2):106-21. Available from: http://dx.doi.org/10.22267/rcia.143102.35 Spanish.

61. Oldfield TL, Sikirica N, Mondini C, Lopez G, Kuikman PJ, Holden NM. Biochar, compost and biochar-compost blend as options to recover nutrients and sequester carbon. J Environ Manag [Internet]. 2018 Jul 15 [cited 2018 Aug];218:465-76. Available from: https://doi.org/10.1016/j.jenvman.2018.04.061 Subscription required to view.

62. Roxas HR, Echeverria HE, Angelini HP. Niveles de carbono orgánico y pH en suelos agrícolas de las regiones pampeana y extrapampeana argentina [Organic carbon and pH levels in agricultural soils of the pampa and extra-pampean regions of Argentina]. Cienc Suelo. 2011 Jul;29(1):29-37. Spanish.

63. Julca-Otiniano A, Meneses-Florian L, Blas-Servillano R, Bello-Amez S. La materia orgánica, importancia y experiencia de su uso en la agricultura [Organic matter, importance, experiences and it role in agriculture]. IDESIA (Arica). 2006;24(1):49-61. Spanish.

64. Beesley L, Moreno E, Fellet G, Carrijo L, Szimur T. Biochar and heavy metals. In: Lehmann J, Joseph S, editors. Biochar for environmental management: science, technology and implementation. 2nd ed. Earthscan: London, UK; 2015. Chapter 20.

65. Tang JJ, Ly H, Gong Y, Huang Y. Preparation and characterization of a novel graphene/biochar composite for aqueous phenanthrene and mercury removal. Bioreour Technol [Internet]. 2015 Nov [cited 2018 Feb];196:355-63 Available from: https://doi.org/10.1016/j.biortech.2015.07.047 Subscription required to view.

66. Thies JE, Rillig MC, Graber E. Biochar effects on the abundance, activity and diversity of the soil biota. In: Lehmann J, Joseph S, editors. Biochar for environmental management: science, technology and implementation. 2nd ed. Earthscan: London, UK; 2015. Chapter 13.

67. Pinedo AU. Obtención de biocarbones y biocombustibles mediante pirolisis de biomasa residual [master's thesis]. [Madrid, Spain]: Universidad Nacional de Educación a Distancia; 2013 Sep. 83 p. Spanish.

68. Informe de ensayo BO1705848. Bogota, Colombia: SGS Colombia S.A.S.; 2017.