Assessment of the Main Natural Disturbances on Norwegian Forest Based on 20 Years of National Inventory

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Abstract

The re-measurement of permanent forest inventories offers a unique opportunity to assess the occurrence and impact of forest disturbances. The present study aims at exploring the main forest damages in Norway based on the extensive data of several consecutive national forest inventories during the period 1995–2014. Five of the most common disturbance agents in Norway are selected for analysis: wind, snow, browsing, fungus and insect damage. The analyses focuses on the frequency and variation along time, the average damage at stand level and the spatial patterns of damage occurrence, resulting in a characterization of the damage produced by disturbances in Norway. The highest damage occurrences by disturbance agent are due to browsing, snow and wind. Snow presents a decreasing temporally trend in damage frequency in the studied period. By forest type, mature and intermediate birch forest are found to be more affected by snow damage, whereas mature spruce forest is by wind damage. The results from this study provide support to the hypothesis that damages by autumnal moth (Epirrita autumnata) on birch are more common in mature stands. No major attacks from bark beetle (Ips typographus) are found, probably related to the lack of major storm damages in the period. Forest types susceptibility to fungus has no apparent variation over time except in the last years, as increased occurrence is observed on mature spruce stands probably correlated with warmer than average periods. Browsing damage causes the most severe losses, as expected, in young stands, and is allocated mainly on the most productive forests. Although some of the disturbances present locally moderate effects, the results show no major disturbances threatening Norwegian forests in the studied period. Finally, the Norwegian national forest inventory demonstrates its reliability as a basis to understand the occurrence and effects of major natural disturbances.
1. Introduction

Natural disturbances are a key factor in forest dynamics [1], and a reason for the potential distribution of terrestrial vegetation [2]. At the same time, disturbances are considered as a major threat to forest resources [3] and associated ecosystem services [4]. The impact either positive or negative, of natural disturbances greatly depends on their regime, and how the affected ecosystems are adapted to cope with the intensity and recurrence of those disturbances. High intensity (i.e. stand level or larger) events are usually attracting most of the attention of society and, arguably, the scientific community. Nevertheless, disturbances have an important impact on the evolution of a forest even when their intensity and associated severity are mild.

Understanding e.g. the establishment of a stand after a severe disturbance, or the diversification of species and structure coming from a frequent but mild disturbance, provides the forester with information that can be included in management plans and objectives. This assertion can be applied either, if the objective is to emulate through management the evolution pathways of a natural forest [5,6], account for expected losses when defining a management plan [7], or combine economic and risk mitigation objectives [8].

In this context, the occurrence, susceptibility and impact of natural disturbances on forest ecosystems has been studied at different temporal and spatial scales, relying on a variety of methodological approaches and data sources [4,9]. In general, disturbance regimes are defined by the frequency of the events, their spatial and temporal distribution and impact on forest [10,11]. Even if the components of a disturbance regime depend largely on the type of disturbance [12], the spatial and compositional characteristics of the forest also has an important role on modifying both the extent and severity of disturbances [13]. The variability on disturbance regimes presents a challenge when large areas and multiple disturbance types are to be analyzed, as the spatial and temporal frames should be large enough to reflect variations on the spatial patterns of the events recurrence. For instance, remote sensing tools have shown their usefulness to capture information on the spatial distribution and impact of different types of disturbances at the required spatial and temporal scales [14–16]. However, remote sensing tools rely on variations on the vegetation structure and vitality, being traditionally better adapted to capture large-scale, stand replacing forest disturbances [17]. In addition, they require from previous information about the type of disturbance causing the observed variation on forest conditions.

Field assessments on forest health, on the other hand, offer the possibility of identifying the disturbance agents, and provide reliable information about the rate of damage, even when it is small or located on non-dominant forest strata. This valuable information comes at a high economic cost if large areas are to be monitored over an extended period of time. National Forest Inventories (NFIs), although not designed for the specific purpose of assessing forest health, are considered a potential source of information that can produce and complement disturbance related information [18,19]. Although NFIs are not spatially continuous, they provide a good approximation at the national scale on the spatial distribution and temporal evolution of forest types and associated goods and services [20]. When measurements of forest health are included on the NFIs design, they become an excellent source of data for assessing the influence of forest characteristics on the occurrence and severity of natural disturbances. Examples on the use of NFIs for assessing occurrence and impact of disturbances can be found for pests or diseases [21–23], fire [24–27], wind and snow [28–31], or game related damage [32,33], among others.

The present study aims at exploring forest disturbances in Norway by their specific agent, with emphasis on Snow, Wind, Browsing, Insect, and Fungus related damage. The main focus is to identify spatial and temporal patterns on the occurrence of forest disturbances based on data available upon request and in accordance with the guidelines of the monitoring program of which they are part.

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from four consecutive measurements of the Norwegian NFI, entailing 1995–2014. Additionally, it provides an overall insight on the occurrence and impact of natural disturbances on the Norwegian forest, analysing the susceptibility of different forest types to be affected by disturbances, and an overall assessment of the damage levels.

2. Materials and Methods

2.1. Data sources

The data for the analysis was based on the Norwegian National Forest Inventory (NFI) collected during the 7th, 8th, 9th and 10th measurements, corresponding to the period 1995–2014. The Norwegian NFI is a systematic permanent inventory, where plots allocated on a 3x3 km grid are measured every 5 years. For each permanent plot, variables were measured at three levels: stand (1000 m² around the plot center), plot and tree level (on a circular plot of 250 m²). Forest damage was measured at stand level. The plots included in the analysis were those with damage measurements available and located in productive forest (i.e. expected yield over 1 m³ ha⁻¹ year⁻¹ over bark [34]). During 1995–2005 damage was not recorded in plots that were in regeneration stage. For plots divided by a stand border (e.g. a plot has one part on forest and another on water, or on two forest types with very different productive parameters), only the forested part was considered or the first division when both parts were forested areas.

In total 34,263 plot measurements were considered for the analysis, 8052, 8423, 8895, 8893 in the 7th, 8th, 9th and 10th NFI respectively; these plots entailing most of the forested parts of the country (Fig 1). The country was divided into regions as defined in the official statistics of forest condition and resources in Norway [34] (Fig 1). These regions are considered to have relatively similar forests, topography and climate (Table 1). The northernmost area of the country, Finnmark, was not included as the measurements in this county have started during the 9th NFI with different grid in non boreal forest areas.

Following the NFI instructions, damage from different disturbance agents were recorded in each plot at each inventory following the criteria: it had an effect in the future economic development of the stand, it compromises the regeneration, or it represents a relevant decrease in the volume production or wood quality [34]. Damage was only recorded when it was detected to have occurred within 5 years prior to the NFI measurement.

In each plot, at least one disturbance agent was associated to the observed damage (Table 2). In the cases that several disturbance agents were reported at the same time in the same plot, the analysis only considered those ranked as the main disturbance agent (i.e. that occasioned the higher damage loss in the stand, as defined in [34]). Also at plot level, damage was quantitatively defined as a loss relative to volume, number of trees or crown loss according to the damage type and disturbance agent (Table 2) [34].

The plots were divided according to their dominant species into four forest types. The criterion for assigning the forest type was the relative abundance of the main species expressed in volume distribution for older forest and crown cover for young forest. Therefore, plots were classified as spruce (spruce > 70%), pine (pine > 70%), birch (birch > 70%) and mixed forests (dominated by any or several of the previous species and with more than 10% of broadleaves). Plots were also divided according to their stand development class in three categories: young, intermediate and mature (corresponding with development class categories I and II, III and IV and V respectively in the Norwegian NFI [34]). The change from young to intermediate development class is mainly defined when the trees’ mean diameter exceeds 10 cm in spruce and pine forest, and 7–8 cm in birch forest.
2.2. Methodological approaches

Five of the most common disturbance agents in Norway were selected for further analysis: wind, snow, browsing, fungus and insects. The analyses focused on the frequency of disturbance occurrence (presence of damage regardless of its impact), its temporal and spatial variation. The variation of the relative frequency of disturbance occurrence along the timeframe of the data was first visually explored by disturbance agent and region. In order to identify possible trends or peaks, linear and polynomial models were developed for each disturbance, identifying those with significant levels (p < 0.05). The level of damage was also assessed quantitatively at plot level, and it was compared between regions.

![Fig 1. Studied area. Left: Studied area and regionalization used in the analysis. The counties were grouped in six main regions, except the northernmost county (Finnmark) that was not included in the analysis. Right: Spatial distribution of the national forest inventory plots by dominant species (spruce, pine, birch and mixed). Map border lines adapted from [35], original licensed under Creative Commons Attribution 4.0 International (CC BY 4.0).](image)

### Table 1. Climatic characterization of the plots included in the analysis, by region.

| Region   | $t_m$ (°C) | $t_{min}$ (°C) | $t_{max}$ (°C) | $P_{sum}$ | N  |
|----------|------------|---------------|---------------|----------|----|
| Region 1 | 3          | -0.7          | 6.7           | 664.7    | 8636|
| Region 2 | 2.4        | -1.5          | 6.3           | 691.2    | 6490|
| Region 3 | 4.9        | 2             | 7.7           | 1049.6   | 5161|
| Region 4 | 5.5        | 2.8           | 8.1           | 1636.6   | 4347|
| Region 5 | 3.4        | 0             | 6.8           | 990.9    | 4889|
| Region 6 | 2.5        | -0.5          | 5.5           | 1040.4   | 4740|

Data based on the averages of the normal climatic period 1960–1990. $t_m$: annual mean temperature, $t_{min}$: minimum temperature, $t_{max}$: maximum temperature, $P_{sum}$: annual precipitation. N: Total number of plots included by region.

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The spatial analysis included the identification of areas with the highest concentration of plots affected by the occurrence of each of the disturbance agents. In this case, a geo-statistical approach based on kernel methods was used [36,37]. Kernel methods allow estimating the probability of occurrence of events in a continuous space. In general, the method requires to define a bandwidth parameter that will be used to aggregate the events. In our case, the events were defined as plots presenting damage by one of the disturbance agents selected. The approach taken was based on an adaptive bandwidth method [38] that varies the bandwidth size in relation with the estimated local amount of data (i.e. the size of the bandwidth is inversely related to the amount of data). Although fixed bandwidths are more common because of its simple and effective application [39,40]; it is widely known that they can present problems when dealing with populations showing high spatial inhomogeneity. A fixed large bandwidth will miss the finer variations of highly dense areas and a narrow one will increase the intensity function for isolated points. The global bandwidth needed to obtain the varying bandwidths was calculated based on the over-smoothing factor [41] and the pilot bandwidth was calculated with a leave-one-out least-squares cross-validation (following Bowman and Azzalini, 1997 [42]). Considering the large border effect that the Norwegian geography can have on the estimates, edge corrected adaptive densities were applied. For the Kernel calculations, the R package `sparr` [43] was used.

| Disturbance agent | Damage                                                                 |
|-------------------|------------------------------------------------------------------------|
| Wind              | Volume of trees blowdown by wind as a percentage of total volume.      |
| Snow              | Number of trees with snow break/blowdown in percentage of the total number of trees. |
| Drought           | Volume of dead trees as a percentage of total volume.                  |
| Frost             | Percentage of stand crown mass that is dead.                          |
| Fire              | Volume of dead trees as a percentage of total volume.                  |
| Landslide         | Number of trees with break / blowdown as a percentage of total number of trees. |
| Browsing          | In old forest, it is the percentage of stand crown mass that is grazed away. In young forest it refers to the percentage of dead or damaged future trees as a percentage of the original number of trees. |
| Insects (includes: Bark beetle (Ips typographus), European pine sawfly (Neodiprion sertifer), Autumnal moth (Epirrita autumnata), Weevil (Curculionidae) and insect not specified) | Percentage of stands crown dead or fell off/ grazed away except for Bark beetle damage that corresponds with volume of dead trees as a percentage of total volume. |
| Fungus (includes: Spruce needle rust fungus (Chrysomyxa abietis), Lophodermium needle cast pine (Lophodermium) and fungus not specified) | Percentage of stand crown that is dead or discoloured for Spruce needle rust fungus and other fungus no specified. And in young forest, percentage of dead future trees as a percentage of the original number of trees. |
| Mechanical (includes: Forest operations or damage by animal) | Volume of trees damaged in percentage of the total volume. |
| Mouse, beaver or ungulates | Volume of damaged trees as a percentage of total volume in old forest. For young forest it is the percentage of dead future trees from original number of trees. |
| Damage cause not known | Percentage of stands total crown that is dead. |

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The frequencies of disturbance occurrence by development class of the forest (i.e. young stands, intermediate and mature) and dominant species were studied using contingency tables. The contingency tables were defined comparing the observed number of plots presenting damage by disturbance agent to the expected frequency values when the number of plots is divided proportionally between development classes. The difference was considered relevant when the observed value per cell was equal or larger than 1.5 the expected value (one-tailed), and otherwise it was assumed no effect of development class by species on the occurrence of the studied disturbance agent. Cells with large difference between observed and expected make a larger contribution to the Chi-squared test. The Chi-squared test was calculated against the null hypothesis that the different disturbance agents have the same effect on all forest types and development classes’ combinations (at the 0.05 level). The characteristics of the stand prior to the damage identification were based on the period before the damage was measured, for example if the disturbance occurrence was measured in the period 2000–2004 (8th NFI) the stand characteristics before the damage corresponds with the information obtained during the period 1995–1999 (7th NFI) in the same plot.

From the initial set of 34 263 plot measurements, all of them were used for temporal damage frequency analysis by disturbance agent. Relative frequencies of damage occurrence by disturbance were calculated relative to the number of plots measured in each year. The Norwegian NFI measures one fifth of all the plots in each year, therefore relative yearly estimations are possible to obtain. Damage was analyzed using the plots affected by one of the selected natural disturbances and with measurements of damage impact on the forest (2557 plots); these measurements cover the period (2000–2014) because prior to 2000 only information on disturbance occurrence (damage or undamaged for each plot) had been recorded. Finally, in the contingency analysis were also included plots with species composition and development class information before the damage (23 767 plots).

### 3. Results

Damage was detected in 3771 plots out of 34 263 studied plots during the period 1995–2014 (about 11%). In relation to the total plots available per forest type, birch forest was the most affected type, followed by mixed and spruce forests (17%, 10% and 8% of all plots, respectively). The highest disturbance occurrences corresponded to: browsing, snow and wind. By forest type, spruce forests were found to be more affected by snow, wind, fungus and browsing damage; pine forest was mainly affected by browsing, fungus and wind, birch forest by snow, insect attacks and browsing, and mixed forest by browsing, snow and wind.

The time series showed some trends in the relative frequency of disturbance occurrences over time (Fig 2). At country level, the occurrence of snow and wind declined over time (linear trend, \( p < 0.001 \) in both cases), although presenting a small increase in 2012–2014. Regionally, snow damage showed a similar trend in each of the 6 the regions (linear trend, \( p < 0.001 \) for all regions), whereas wind showed a decreasing trend only in some regions (linear trend, \( p < 0.001 \) only for regions 4, 5 and 6) as well as browsing (linear trend, \( p = 0.02 \) for region 6). Curves with peaks of frequencies were observed for browsing (significant for regions 3 and 5), and fungus (significant for region 6). In the case of snow and insect damage there were large occurrence pikes in region 6, corresponding to the years 1997, 2003 and 2013.

Each disturbance type showed a specific spatial distribution (Fig 3). Snow related damages were mainly located in the southern regions (region 1 and 2) and in the northern region (region 6), being specially clustered in mountainous areas. Wind related damage was more frequent on the western regions (region 4 and 5) and in mountainous areas facing the Atlantic side. Browsing damage was clearly aggregated in southeast regions (regions 1–3), forming two
Fig 2. Frequencies of damage occurrence by disturbance agent along 1995–2014. Top: Relative to country level (All). Bottom: Relative to each region.

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main hotspots. Insect related damage was most common in the northernmost of region 6, forming disperse hotspots. The occurrence of fungal damage was mainly present in the south of the country, with three main hotspots: in the coast, at the Swedish border and in the mountain areas of region 2.

It must be taken into account that damage due to insect and fungus responds to different fungal agents that are specific to a species, and the maps represent only the general occurrence of fungal damages altogether. Therefore, the contingency tables aimed at observing the damage
by both species composition and development classes (Table 3), in order to identify difference occurrence rates for different species or for different development stages related to the stand’s age. The calculated Chi-squared was 4117.6 (d.f. = 55) with a p-value < 0.001, presenting a highly significant departure from the null hypothesis, meaning that the different damage agents affect differently to different forest types and development classes. For instance, the observed frequency for snow damage was considerably higher from expected values in mature and intermediate birch forests. The occurrence of wind damage was higher than expected in mature spruce forest (regions 2, 3 and 5). The occurrence of browsing damage was higher than expected in young stands of all forest types except in spruce stands. In spruce stands, all stages of development appeared to be less susceptible to browsing damage, except during the period 2005–09 where young spruce stands followed the same trend as other young forest types. Mature and intermediate birch forest were more susceptible to insect damage than other forest types. None of the forest types was found more susceptible of fungus damages over the whole study period. Although, during the last years spruce seemed to be more prone to suffer from fungus attacks. The effect of each of the disturbances agents, on the different forest types, was also tested showing that individually, the damage agents also affect differently to different forest types and development classes. The Chi-squared results were 458.46 for snow, 222.82 for wind, 2339 for browsing, 726.57 for insect and 70.44 for fungus, (d.f. = 11 and p-value < 0.001 for all cases).

Finally, when the level of damage per affected plot was evaluated, it was found that the disturbance agent causing more severe losses was browsing, and especially on those plots located in regions 1–4 (Fig 4). The mean percentage of damage losses caused by snow, wind and fungus damages were not significantly different between regions. Most of the losses due to insect damage were in region 6 where the species composition is typically dominated by birch.

4. Discussion

Long-term damage analysis enhances the understanding of natural disturbances in forest areas [30] and it can help forecast the probability of damage given different timeframes. Previous studies have demonstrated the reliability of NFI data to assess the occurrence of the most

Table 3. Contingency table presenting observed and calculated expected stand damage by disturbance agent.

| Development class | Spp. Composition | Snow | Wind | Browsing | Insect | Fungus | No damage |
|-------------------|------------------|-------|------|----------|--------|---------|-----------|
| Mature Spruce     | 46 / 24          | 55 / 22 | 13 / 24 | 26 / 23 | 37 / 27 | 1975 / 2016 |
| Pine              | 5 / 23           | 18 / 23 | 11 / 24 | 1 / 22  | 33 / 27 | 2447 / 2206 |
| Birch             | 132 / 24         | 25 / 24 | 6 / 24 | 132 / 24 | 7 / 25 | 1151 / 1053 |
| Mixed Spruce      | 44 / 24          | 14 / 24 | 9 / 24 | 15 / 24 | 14 / 25 | 1947 / 1052 |
| Pine              | 3 / 23           | 16 / 23 | 8 / 23 | 5 / 23 | 37 / 23 | 2690 / 1997 |
| Birch             | 22 / 27          | 6 / 24 | 8 / 24 | 89 / 24 | 11 / 23 | 1150 / 1277 |
| Mixed Spruce      | 42 / 24          | 26 / 24 | 16 / 24 | 13 / 24 | 37 / 24 | 3223 / 2080 |
| Pine              | 9 / 15           | 1 / 15 | 0 / 15 | 0 / 15 | 29 / 15 | 1152 / 1177 |
| Birch             | 4 / 13           | 0 / 13 | 11 / 13 | 8 / 13 | 89 / 13 | 1157 / 1058 |
| Mixed Spruce      | 4 / 15           | 0 / 15 | 11 / 15 | 8 / 15 | 89 / 15 | 1152 / 1080 |
| Pine              | 9 / 23           | 0 / 23 | 420 / 23 | 5 / 23 | 29 / 23 | 2233 / 2830 |

Highlighted, those presenting a relevant deviation between observed and expected.

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important disturbances and their impact on forest resources [21,23,44], making NFI data a good basis for a characterization of disturbances in the forest. Variables from NFI have also been used to better understand forest factors related to occurrence or damage of different natural disturbances [24,28,32].

The present study evaluates the temporal evolution and spatial aggregation patterns of disturbance occurrence across Norway, based on several consecutive NFI datasets entailing all productive forest and showing its state and development (stock, species composition, health) over a 20 year period. The availability of four NFI measurements provides an excellent opportunity to evaluate similarities and divergences between disturbances, considering their temporal and spatial aggregation patterns.

When evaluating the evolution of disturbance occurrence, we observed clear variations between disturbance agents. The occurrence of snow and wind related damages appeared to peak in different years, but snow had a much higher peak between 1996–1997, and lower ones after this period. It was not possible to identify a clear fixed interval, which agrees with other European records where snow damaging events varied from each winter to longer cycles [45]. As expected, location seems to play a major role on defining the susceptibility to suffer from snow and wind damage [46]. As many other mountainous countries, Norway, presents a regionalized climate that can influence the spatial distribution of affected plots. Southern areas of the country are often influenced by southern weather systems whereas northern areas are influenced by oceanic clime, due to the proximity of the north Atlantic streams on the west. We found that the occurrence of snow and wind damages tended to be aggregated in mountainous areas, with no major differences along the north-south axis. Typically, northern trees are more adapted to this type of damage and also have a slower growth, allowing them to better adapt to wind and snow damage events [28].

When defining which forest types were more susceptible to be affected by wind and snow, our analysis showed that snow damage was more frequent on mature and intermediate birch forest, whereas wind affectation was associated to mature spruce, and both wind and snow damages were rare on pine forest. One factor that can explain this is the higher exposure of birch stands to the snow due to their location in mountainous areas. Another factor relates with structural parameters, e.g.: lower spacing between trees, compared to conifers. Our results partially disagree with previous studies indicating that snow and wind damage are more
characterisation of natural disturbances in norway

common in scots pine and norway spruce than birch [47,48]. however, for a similar diameter class, scots pine would be more resistant to uprooting than spruce and birch stands, due to its deeper roots and better anchorage [47]. in addition to higher exposures due to location, one potential reason for the susceptibility of birch to snow damage could be that in norway birch forest is usually shorter in height than spruce forest and their stems are easily bent or blow down by snow load or avalanches. birch also has a tendency to regenerate through coppice in small groups of stems, generating gaps between regeneration groups that can ease the appearance of snow damage, for example due to avalanches [49]. in the case of spruce and wind damage, the higher susceptibility of the older stages of development was consistent with some studies [50] but disagreed with others [51]. it has to be mentioned that as the nfi data on wind damage focus on recording blown down trees, uprooting processes are overlooked in contrast to possible steam breakage, which is allocated as snow damage. these criteria in data recording can further explain our results, as larger specimens of spruce are more susceptible to uprooting [52] in contrast to smaller trees that often are more prone to break.

in the case of biotic agents, it is important to mention that the norwegian nfi focused on some specific agents, both in case of insects and pathogens. for instance, in the case of insects, priority was given to bark beetle (ips typographus), european pine sawfly (neodiprion sertifer), autumal moth (epirrita autumnata) and weevil (curculionidae), each of them typically specialized in certain tree species. the joined analysis of different and specialized insect and fungus agents had an important effect on all the dimensions of the occurrence analysis, and was therefore further explored by species in the contingency analysis. for example, insect damage was especially visible in a region dominated by birch (region 6), as most of the insect related damage corresponds to autumal moth attacks on birch. the temporal evolution of the attacks of insect frequency showed cyclic occurrence peaks and agrees with previous studies where autumal moth typically present cyclical outbreaks every 10 years [53,54]. our results also provided support to the hypothesis that autumal moth on birch is more common in older stands [55,56], as mature birch stands offer more favorable places for oviposition and an enhancement of food resources [55]. another interesting result is the lack of major bark beetle outbreaks on spruce forest, during the period 1995–2014. this result is probably associated to the limited storm related damage for the same period and the well-known interaction between storm damage and ips outbreaks [57–59]. the observed trend of limited but highly variable presence of bark beetle damage, agrees with the trends identified for the period 1972–2002 by økland and bjønstad (2003) [60], when no important outbreak was detected the years following the major attack of the 70s [61,62].

concerning fungal attacks, the records also corresponded with different and specialised agents: spruce needle rust fungus (chrysomyxa abietis) on spruce and lophodermium needle cast pine (lophodermium) on pine. the selection of agents had a clear effect on the allocation patterns of the recorded attacks, as the most affected regions corresponded to those were spruce was the dominant tree species. regarding the temporal evolution of the fungi attacks, they correlated with humid springs [63], and in many cases warmer than average periods such as the 2002–2004 and specially 2013–2014. when the susceptibility of forest types to fungi was analysed, no apparent variations were identified but during the last years analysed, where an increase on attack susceptibility was observed on young or mature spruce stands (s1 table). browsing was one of the most relevant disturbance agents associated to norwegian forests, both on occurrence and impact on the forest. browsing frequency remained similar across time and the damage was allocated mainly on the most productive forests in the south of the country. moose habitat selection is led by forage quality and shelter availability, both aspects changing in space and time [64]. browsing affected mainly young stands, as typically moose prefer those because they offer more palatable forage at a reachable height [32]. the species

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composition of the stands also had an important effect on the allocation of browsing damage, due to specific food choices and preferences by ungulates [9,32]. Our study shows the preference by ungulates of mixed, birch and pine stands, except for the period 2005–2009 when spruce was also significantly damaged (S1 Table). Typically, moose prefer spruce stands the least [32], although it is also known that increased densities of moose population might also increase the risk of browsing damage (e.g. stands on migration routes or next to winter habitats). An important data limitation was that the 7th and 8th NFI (1995–2005) did not include damage on forests in regeneration stage. It would be reasonable to assume that the occurrence of browsing on those young stands would be even higher, as moose is known to browse in mixed stands with young pines (age<5 years) [65].

Despite the limitations, the Norwegian NFI has shown its potential as data source for assessing natural disturbances, their spatial and temporal distribution and the vulnerability of different forest types to them. We consider that the combination of four different NFI measurements provides a unique opportunity for this type of analysis. Our analysis suggests that a clear relation exists between wind and snow damage, which is supported by the recognized combined effect of wind intensity and snow load to induce storm related damage [28,51,66]. However, the instructions of the Norwegian NFI lead to consider stem breakage as snow derived damage (and never as a wind related damage) which could induce to misinterpretations. In this study we did not consider interactions among different disturbance agents due to the complexity of the analysis required and because of the absence of major abiotic disturbances that might trigger potential future damage by biotic agents [49,67–69]. The combination of different disturbance types within the same measurement was not considered, as the main disturbance was the only registered damage in over 85% of the damaged plots (S2 Table).

Finally, not finding major disturbances threatening Norwegian forests is by itself the main result of the study. After the catastrophic storms and bark beetle attacks of the 70s [62,70], we may say that the Norwegian forest has sustained a relatively long period of good health, either for the absence of extreme storms, or due to the effects that past storms and biotic attacks had on the temporal depletion of susceptible trees [71]. This assumption calls for further analysis such as predictions of the forest evolution to detect potential threats in the future due to the natural ageing of the forest, or the inclusion of structural or management dependent variables, aiming at identifying past management actions that may help mitigate the impact of each disturbance. For example, the inclusion of more detailed information about snow loads and stand management history would also help define the importance of variations in the stand and trees’ structure on storm related damage.

The study of forest disturbances is an important source of information for developing a holistic management of forests and its associated ecosystem services [4]. The analysis of records of damage by different causing agents, with spatial approaches or with descriptors of forest stands can therefore help identify susceptible forest areas and at the same time, forest management practices that can reduce their probability.

Supporting Information

S1 Table. Contingency tables. Tables presenting observed (numerator) and the calculated expected (denominator) stand damage by development class and disturbance agent and Norwegian National Forest Inventory (NFI).

(DOCX)

S2 Table. Disturbance agents simultaneously recorded in the same measurement. Percentage of the plots damaged by the main disturbance agent that were simultaneously damaged by
a secondary and tertiary disturbance agent.

Author Contributions

Conceptualization: ODY BMY JGO.
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Writing – original draft: ODY BMY JGO.
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References

1. Oliver CD, Larson BC. Forest stand dynamics. 1st ed. New York; 1996.
2. Bond WJ, Keeley JE. Fire as a global “herbivore”: the ecology and evolution of flammable ecosystems. Trends in Ecology & Evolution. Elsevier; 2005; 20: 387–394. doi:10.1016/j.tree.2005.04.025
3. Schelhaas MJ, Nabuurs GJ, Schuck A. Natural disturbances in the European forests in the 19th and 20th centuries. Global Change Biology. 2003; 9: 1620–1633. doi:10.1046/j.1365-2486.2003.00684.x
4. Thom D, Seidl R. Natural disturbance impacts on ecosystem services and biodiversity in temperate and boreal forests. Biol Rev. 2015. doi:10.1111/brv.12193
5. Bergeron Y, Gauthier S, Kafka V, Lefort P, Lesieur D. Natural fire frequency for the eastern Canadian boreal forest: consequences for sustainable forestry. Can J For Res. 2001; 31: 384–391. doi:10.1139/cjfr-31-3-384
6. Franklin JF, Spies TA, Van Pelt R, Carey AB, Thornburgh DA, Berg DR, et al. Disturbances and structural development of natural forest ecosystems with silvicultural implications, using Douglas-fir forests as an example. Forest Ecology and Management. 2002; 155: 399–423. doi:10.1016/S0378-1127(01)00575-8
7. Gadow von K. Evaluating risk in forest planning models. Silva Fennica. 2000; 34: 181–191.
8. González-Olabarria JR, Pukkala T. Integrating fire risk considerations in landscape-level forest planning. Forest Ecology and Management. 2011; 261: 278–287. doi:10.1016/j.foreco.2010.10.017
9. Seidl R, Fernandes PM, Fonseca TF, Gillet F, Jönsson AM, Merganičová K, et al. Modelling natural disturbances in forest ecosystems: a review. Ecological Modelling.Elsevier B.V; 2011; 222: 903–924. doi:10.1016/j.ecolmodel.2010.09.040
10. Johnson EA, Gutsell SL. Fire frequency models, methods and interpretations. Advances in Ecological Research. Elsevier; 1994; 25: 239–287. doi:10.1016/S0065-2504(08)60216-0
11. Runkle JR. Disturbance regimes in temperate forests. In: Pickett STA, White PS, editors. The ecology of natural disturbance and patch dynamics. London; 1985. pp. 17–33.
12. Fischer A, Marshall P, Camp A. Disturbances in deciduous temperate forest ecosystems of the northern hemisphere: their effects on both recent and future forest development. Biodivers Conserv. Springer Netherlands; 2013; 22: 1863–1893. doi:10.1007/s10531-013-0525-1
13. Bond WJ, Midgley JJ. Kill thy neighbour: an individualistic argument for the evolution of flammability. Oikos. 1995; 73: 79. doi:10.2307/3545728

14. Hall FG, Botkin DB, Strebel DE, Woods KD, Goetz SJ. Large-scale patterns of forest succession as determined by remote sensing. Ecology. 1991; 72: 628–640.

15. Franklin J. Predictive vegetation mapping: geographic modelling of biospatial patterns in relation to environmental gradients. Progress in Physical Geography. SAGE Publications; 1995; 19: 474–499. doi:10.1177/030913339501900403

16. Chuvieco E. Remote sensing of large wildfires. Chuvieco E, editor. Berlin, Heidelberg: Springer Science & Business Media; 2012. doi:10.1007/978-3-642-60164-4

17. Frolking S, Palace MW, Clark DB, Chambers JQ, Shugart HH, Hurtt GC. Forest disturbance and recovery: A general review in the context of spaceborne remote sensing of impacts on aboveground biomass and canopy structure. J Geophys Res. 2009; 114: n/a–n/a. doi:10.1029/2008JG000911

18. Allen E. Forest health assessment in Canada. Ecosystem Health. 2001; 7: 28–34.

19. Wulff S, Roberge C, Ringvall A, Holm S, Ståhl G. On the possibility to monitor and assess forest damage within large scale monitoring programmes—a simulation study. Silva Fennica. 2013;47. doi:10.14214/sf.1000

20. Bergstedt J, Milberg P. The impact of logging intensity on field-layer vegetation in Swedish boreal forests. Forest Ecology and Management. 2001; 154: 105–115. doi:10.1016/S0378-1127(00)00642-3

21. Nevalainen S. Gremmeniella abietina in Finnish Pinus sylvestris stands in 1986–1992: A study based on the national forest inventory. Scandinavian Journal of Forest Research. 1999; 14: 111–120. doi: 10.1080/02827589950152836

22. Nevalainen S, Yli-Kojola H. Extent of abiotic damage and its relation to defoliation of conifers in Finland. Forest Ecology and Management. 2000; 135: 229–235. doi:10.1016/S0378-1127(00)00313-3

23. Nevalainen S, Sirkiä S, Peltoniemi M, Neuvonen S. Vulnerability to pine sawfly damage decreases with site fertility but the opposite is true with Scleroderris canker damage; results from Finnish ICP Forests and NFI data. Ann For Sci. 2015; 72: 909–917. doi: 10.1007/s13595-014-0435-8

24. González JR, Palahí M, Trasobares A, Pukkala T. A fire probability model for forest stands in Catalonia (north-east Spain). Ann For Sci. 2006; 63: 169–176. doi: 10.1051/forest:2005109

25. González JR, Trasobares A, Palahí M, Pukkala T. Predicting stand damage and tree survival in burned forests in Catalonia (North-East Spain). Ann For Sci. 2007; 64: 733–742. doi: 10.1051/forest:2007053

26. García-Gonzalo J, Pukkala T, Borges JG. Integrating fire risk in stand management scheduling. An application to Maritime pine stands in Portugal. Ann Oper Res. Springer US; 2011; 219: 379–395. doi: 10.1007/s10479-011-0908-1

27. Marques S, García-Gonzalo J, Botequim B, Ricardo A, Borges JG, Tome M, et al. Assessing wild fire occurrence probability in Pinus pinaster Ait. stands in Portugal. Forest Syst. 2012; 21: 111–120.

28. Valfinger E, Fridman J. Models to assess the risk of snow and wind damage in pine, spruce, and birch forests in Sweden. Environmental Management. 1999; 24: 209–217. PMID:10384030

29. Dobbertin M. Influence of stand structure and site factors on wind damage comparing the storms Vivian and Lothar. For Snow Landsc Res. 2002; 77: 187–205.

30. Hanewinkel M, Breidenbach J, Neeff T, Kublin E. Seventy-seven years of natural disturbances in a mountain forest area—the influence of storm, snow, and insect damage analysed with a long-term time series. Can J For Res. 2008; 38: 2249–2261. doi: 10.1139/X08-070

31. Martín-Alcón S, González-Olabarria JR, Coll L. Wind and snow damage in the Pyrenees pine forests: effect of stand attributes and location. Silva Fennica. 2010; 44: 399–410.

32. Jalkanen A. The probability of moose damage at the stand level in southern Finland. Silva Fennica. 2001; 35: 159–168.

33. Vospernik S. Probability of bark stripping damage by red deer (Cervus elaphus) in Austria. Silva Fennica. 2006; 40: 589–601.

34. Landsskogtakseringens. Landsskogtakseringens feltinstruks 2008. 2008; 1–153.

35. Kartverket. N250, Norway; 2015. Available: www.kartverket.no.

36. Warton BJ. Kernel methods for estimating the utilization distribution in home-range studies. Ecology. 1989; 70: 164. doi: 10.2307/1938423

37. Seaman DE, Powell RA. An evaluation of the accuracy of kernel density estimators for home range analysis. Ecology. 1996; 77: 2075. doi: 10.2307/2265701

38. Abramson IS. On bandwidth variation in kernel estimates—a square root law. The Annals of Statistics. 1982.
39. Koutsias N, Kalabokidis KD, Allgöwer B. Fire occurrence patterns at landscape level: Beyond positional accuracy of ignition points with kernel density estimation methods. Natural Resource Modeling. 2008; 17: 359–375. doi: 10.1111/j.1939-7445.2004.tb00141.x

40. Ia Riva de J, Pérez-Cabello F, Lana-Renault N, Koutsias N. Mapping wildfire occurrence at regional scale. Remote Sensing of Environment. 2004; 92: 288–294. doi: 10.1016/j.rse.2004.06.013

41. Terrell GR. The maximal smoothing principle in density estimation. Journal of the American Statistical Association. 1990; 85: 470–477. doi: 10.1080/01621459.1990.10476223

42. Bowman AW, Azzalini A. Applied smoothing techniques for data analysis: The Kernel approach with S-Plus illustrations. Oxford; 1997.

43. Davies TM, Hazelton ML, Marshall JC. sparr: Analyzing Spatial Relative Risk Using Fixed and Adaptive Kernel Density Estimation in R. Journal of Statistical Software. 2011; 39: 1–14. doi: 10.18637/jss.v039.i01

44. Wulff S, Hansson P, Witzell J. The applicability of national forest inventories for estimating forest damage outbreaks—Experiences from a Gremmeniella outbreak in Sweden. Can J For Res. 2006; 36: 2605–2613. doi: 10.1139/x06-148

45. Nykänen ML, Peitola H, Quine C, Kellomäki S. Factors affecting snow damage of trees with particular reference to European conditions. Silva Fennica. 1997; 31: 192–213.

46. Valinger E, Pettersson N. Wind and snow damage in a thinning and fertilization experiment in Picea abies in southern Sweden. Forestry. 1996; 69: 25–33.

47. Peitola H, Nykänen ML, Kellomäki S. Model computations on the critical combination of snow loading and windspeed for snow damage of Scots pine, Norway spruce and birch sp. at stand edge. Forest Ecology and Management. 1997. doi: 10.1016/S0378-1127(97)00037-6

48. Jalkanen R, Konopczka B. Snow-packing as a potential harmful factor on Picea abies, Pinus sylvestris and Betula pubescens at high altitude in northern Finland. European Journal of Forest Pathology, Blackwell Publishing Ltd; 1998; 28: 373–382. doi: 10.1111/j.1439-0329.1998.tb01191.x

49. Bebi P, Kulakowski D, Rixen C. Snow avalanche disturbances in forest ecosystems—State of research and implications for management. Forest Ecology and Management. Elsevier; 2009; 257: 1883–1892.

50. Jalkanen A, Mattila U. Logistic regression models for wind and snow damage in northern Finland based on the National Forest Inventory data. Forest Ecology and Management. Elsevier; 2000; 135: 315–330.

51. Päätalo M-L. Risk of Snow Damage in Unmanaged and Managed Stands of Scots Pine, Norway Spruce and Birch. Scandinavian Journal of Forest Research. 2000; 15: 530–541. doi: 10.1080/028275800750173474

52. Illisson T, Metslaid M, Vodde F, Jõgiste K, Kurm M. Storm disturbance in forest ecosystems in Estonia. Scandinavian Journal of Forest Research. 2005; 20: 88–93. doi: 10.1080/14004080510041020

53. Tenow O, Nilsson AC, Bylund H, Hogstad O. Waves and synchrony in Epirrita autumnata /Operophtera brumata outbreaks. I. Lagged synchrony: regionally, locally and among species. J Anim Ecology. 2007; 76: 258–268. doi: 10.1111/j.1365-2656.2006.01204.x

54. Ruohomäki K, Tanhuanpaa M, Ayres MP, Kaitaniemi P, Tammaru T, Haukojo E. Causes of cyclicity of Epirrita autumnata (Lepidoptera, Geometridae): grandiose theory and tedious practice. Population Ecology. 2000; 42: 211–223.

55. Bylund H. Stand age-structure influence in a low population peak of Epirrita autumnata in a mountain birch forest. 1997;: 319–326.

56. Ruohomäki K, Virtanen T, Kaitaniemi P, Tammaru T. Old Mountain Birches at High Altitudes Are Prone to Outbreaks of Epirrita autumnata (Lepidoptera: Geometridae). Population Ecology. 1997; 26: 1096–1104.

57. Økland B, Bjørnstad ON. A resource-depletion model of forest insect outbreaks. Ecology. 2006; 87: 283–290. doi: 10.18637/jss.v039.i01

58. Stadelmann G, Bugmann H, Wermelinger B, Meier F, Bigler C. A predictive framework to assess spatio-temporal variability of infestations by the European spruce bark beetle. Ecography. Blackwell Publishing Ltd; 2013; 36: 1208–1217. doi: 10.1111/j.1600-0587.2013.01777.x

59. Stadelmann G, Bugmann H, Wermelinger B, Bigler C. Spatial interactions between storm damage and subsequent infestations by the European spruce bark beetle. Forest Ecology and Management. Elsevier; 2014; 318: 167–174. doi: 10.1016/j.foreco.2014.01.022

60. Økland B, Bjørnstad ON. Synchrony and geographical variation of the spruce bark beetle (Ips typographus) during a non-epidemic period. Population Ecology. 2003; 45: 213–219. doi: 10.1007/s10144-003-0157-5

61. Bakke A. Host tree and bark beetle interaction during a mass outbreak of Ips typographus in Norway. Zeitschrift für angewandte Entomologie. 1983. doi: 10.1111/j.1439-0418.1983.tb03651.x
62. Bakke A. The recent Ips typographus outbreak in Norway: Experiences from a control program. Holarctic ecology. 1989.

63. Collins MA. Periodicity of spore liberation in Chrysomyxa abietis. Transactions of the British Mycological Society. British Mycological Society; 1976; 67: 336–339. doi: 10.1016/S0007-1536(76)80143-X

64. Bjørneraas K, Solberg EJ, Herfindal I, Moorter BV, Rolandsen CM, Tremblay J-P, et al. Moose Alces alces habitat use at multiple temporal scales in a human-altered landscape. Wildlife Biology. 2011; 17: 44–54. doi: 10.2981/10-073

65. Heikkilä R, Härkönen S. Moose browsing in young Scots pine stands in relation to forest management. Forest Ecology and Management. 1996; 88: 179–186. doi: 10.1016/S0378-1127(96)03823-6

66. Fridman J, Valinger E. Modelling probability of snow and wind damage using tree, stand, and site characteristics from Pinus sylvestris sample plots. Scandinavian Journal of Forest Research. 1998; 13: 348–356. doi: 10.1080/02827589809382994

67. Netherer S, Nopp-Mayr U. Predisposition assessment systems (PAS) as supportive tools in forest management—rating of site and stand-related hazards of bark beetle infestation in the High Tatra Mountains as an example for system application and verification. Forest Ecology and Management. 2005; 207: 99–107. doi: 10.1016/j.foreco.2004.10.020

68. Christiansen E, Bakke A. Does drought really enhance Ips typographus epidemics? A Scandinavian perspective. In: Gregoire JC, Liebhold AM, Stephen FM, Day KR, Salom SM, editors. 1997. pp. 163–171.

69. Wermelinger B. Ecology and management of the spruce bark beetle Ips typographus—a review of recent research. Forest Ecology and Management. 2004; 202: 67–82. doi: 10.1016/j.foreco.2004.07.018

70. Økland B, Berryman A. Resource dynamic plays a key role in regional fluctuations of the spruce bark beetles Ips typographus. 2004; 6: 141–146.

71. Wermelinger B, Obrist MK, Baur H, Jakoby O, Duelli P. Synchronous rise and fall of bark beetle and parasitoid populations in windthrow areas. Agr Forest Entomol. Blackwell Publishing Ltd; 2013; 15: 301–309. doi: 10.1111/afe.12018