Impact of dry sludges and sludge biochar on height and dry matter of Solanum lycopersicum L.

Impacto de lodos secos y biocarbón de lodos sobre la altura y materia seca de Solanum lycopersicum L.

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ABSTRACT

The generation of sludge as anthropic waste is a fundamental pollution problem. However, its conversion to biochar can be an alternative to conventional fertilization for its management and use in agriculture. In this research, we evaluated the effect of the application of different doses of dry sludge (DS) and biochar of pyrolyzed sludge (PS) on the height and dry matter of a tomato (Solanum lycopersicum L.) crop and the nutrient content in the substrate. The biochar was made by rapid pyrolysis, and the substrate and the dry matter of plants were analyzed by different physical and chemical methods. An evaluation of 11 treatments was carried out in allometric measurements of plants and foliar dry matter, in three replicates with two materials (DS and PS) added to the substrate at different levels. The plant height and dry weight were evaluated using an incomplete factorial design in a completely randomized arrangement by performing statistical analysis of multivariate variance. An increase in plant height and dry weight was observed when the doses of DS and PS were increased; however, there were no statistical differences between the two materials. The amount of carbon, organic matter, and Ca concentrations in the dry leaf weight were increased with the addition of DS and PS. Likewise, the use of these materials as conditioners or amendments to agricultural soil at doses of 10-15 t ha⁻¹ may be viable and can contribute to reducing environmental externalities through the use of these anthropic waste materials.

Key words: substrate, tomato, pyrolysis, soil fertility.

Introduction

The accelerated process in urbanization and industry has generated an increase in the production of waste from industrial, domestic, and agricultural activities. Simultaneously to this development, protection policies have been established for the different ecosystems. However, the associated problems persist, particularly in water ecosystems. The construction of alternatives that mitigate pollution, such as aqueducts, sewers, and industrial and municipal treatment plants (wastewater treatment plants (WWTP)) has been encouraged. Although these plants manage to control water pollution problems to a large extent, some by-products, such as sludge, are generated.

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La generación de lodos como residuos antrópicos es un problema fundamental de contaminación. Sin embargo, su conversión a biocarbón puede ser una alternativa a la fertilización convencional para su manejo y uso en la agricultura. En este trabajo se evaluó el efecto de la aplicación de diferentes dosis de lodos secos (LS) y biocarbón de lodos pirolizados (LP) sobre la altura y materia seca de un cultivo de tomate (Solanum lycopersicum L.) y sobre el contenido de nutrientes en el sustrato. El biocarbón se elaboró por pirólisis rápida, y el sustrato y la materia seca de plantas se analizaron mediante diferentes métodos físicos y químicos. Se realizaron 11 tratamientos en medidas alométricas de plantas, sustratos y materia seca foliar, en tres réplicas con dos materiales (LS y LP) adicionados en diferentes niveles al sustrato. La altura y materia seca de la planta se evaluaron bajo un diseño factorial incompleto en arreglo completamente al azar, realizando un análisis estadístico de varianza multivariante. Se observó un incremento en la altura y materia seca en las plantas cuando aumentaron las dosis de LS y LP; sin embargo, no existieron diferencias estadísticas entre los dos materiales evaluados. La cantidad de carbono, la materia orgánica y las concentraciones de Ca en el sustrato aumentaron con la adición de LS y LP. Así mismo, el uso de estos materiales como acondicionadores o enmiendas del suelo agrícola puede ser viable en dosis de 10-15 t ha⁻¹ aportando en la disminución de externalidades ambientales mediante el uso de estos materiales de desecho antrópico.

Palabras clave: sustrato, tomate, pirólisis, fertilidad del suelo.
These by-products are produced from the accumulation of solids found in the effluent (primary sludge) or from the suspension of solids (activated sludge) that results from dissolved solids in wastewater (Morales, 2005).

For the first decade of this century, the average sludge production worldwide was estimated at approximately 30 kg per person/year (Hospido et al., 2010). In Colombia, approximately 247 t of these sludges are generated (94 t dry weight), of which 97% are produced in three treatment plants: Cañaveralejo, San Fernando, and Salitre (PTAR, 2009). Specifically, Salitre treats wastewater from the Salitre river and the Torca and Conejera wetlands, generating about 150 t/d of sludge that could be used in the reforestation of degraded soils (EAAB, 2009). These sludges are rich in macronutrients such as N and P, micronutrients such as Zn and Mo, and organic matter. However, they also contain some pathogens and heavy metals (Hartman et al., 2003); in consequence, the disposition of these materials must be done carefully, both for the volumes that are generated and for the environmental risks they represent.

An alternative for improving fertility indicators and the quality of degraded soils is the inclusion of exogenous organic materials. Through the application of organic amendments, the nutrient status in the soil is improved, and these amendments can serve as a source of macro and microelements to improve the functions and physical-chemical processes of the soil (Lal, 2016). So, the use of sludge in the soil or as an agricultural substrate can serve as an option for increasing yields of different crops. However, it should be noted that many of these materials must be stabilized before application because of potentially toxic elements and pathogens. Through the composting process, it is possible to obtain dry sludge (DS) for uniform and biologically stable products that can act as a source of nutrients for plants (Sullivan et al., 2002).

One of the strategies reported in the literature for improving soils and reducing or neutralizing these adverse elements is the incineration of these materials through pyrolysis, which consists of the burning of organic materials in the absence or low levels of oxygen. One of the products obtained under this process is biochar, also known as pyrolyses or biocarbon. These are solid products of fine and porous grain obtained from the thermal conversion of biomass in a limited oxygen environment. Biochar is similar to coal, which is produced by natural burning, and that reaches great stability in the soil compared to the nitric Nitrogen from which it comes (Lehmann and Joseph, 2009; IBI, 2012). According to Amonette (2009), biochar has a high content of organic carbon, with high resistance to decomposition and high residence time in soils. These characteristics prevent the transformation of biochar and its early release of CO₂ into the atmosphere. Biochar creates a recalcitrant carbon deposit (carbon-negative) that functions as a carbon sequestration network (FAO, 2004).

Both the direct application of dry sludge (DS) in agricultural systems and the conversion of DS to pyrolyzed sludge (PS) for later incorporation can represent an economically and environmentally viable alternative for the disposal of these materials reducing their polluting potential. Kistler et al. (1987), Paterson et al. (2008), and Hossain et al. (2010) showed that the addition of biochar from sludge to tomato crops increased the agrological properties of the soil, which increased crop yield and production. Likewise, these authors claim that the heavy metals concentrated in the solid product become more chemically and biologically inert, decreasing transport or bioaccumulation in the plant and in soils when biochar is used as an amendment.

Nzanza et al. (2012) report that no benefits were found for fruit yield and nutrient absorption with the addition of PS and tree mycorrhizae to the soil of a tomato crop. Because of contrasting results like these found in the literature, more research is needed for plant species of great agricultural importance regarding the effect of biochar application to the soil/substrate, particularly when DS is used.
Materials and methods

Analysis of soil samples
The soil used in the experiment was a degraded soil from the municipality of La Vega (Colombia). The soil samples were used as a substrate and were placed in each of the experimental units: pots of 15.5 cm for the upper diameter, 11.5 cm for the lower diameter, with heights of 11.0 cm and two-liter capacity. Previously, the samples were dried at room temperature for 10 d and then placed in a REDLINE RF 115 mechanical convection drying oven (General Lab & Cleanroom Supply, California, USA) at 70°C for 72 h. The samples were then ground and passed through a 2 mm sieve and sent to the Soil and water laboratory at Universidad Jorge Tadeo Lozano, Bogota, Colombia, for analysis.

We quantified organic matter according to the methodology of Walkley and Black (Kumada, 1987) and SSL (Soil Survey Laboratory) (1995). N was determined using Kjeldahl digestion. We measured the pH in water in a 1:1 ratio (You et al., 1999) in 20 g of soil/20 g of H₂O. The CEC was determined based on the sum of the bases, extracted with NH₄O Ac 1 N at pH 7 (BT). We determined the exchangeable acidity (EA) with KCI 1 N (Motta, 1990) and phosphorus availability with lactate (Egner, 1941). The B was analyzed with azomectin H (Wolf, 1974), the S was determined with turbidimetry (Combatt et al., 2014), and the Fe, Mn, Cu and Zn were analyzed using the diethylene triamine-pentacetic acid (DTPA) method.

Foliar analysis in the different treatments
Destructive sampling was carried out using six-month-old plants, which were dried at 70°C for 72 h and then incinerated. The incinerated material was analyzed to determine different macro and microelement values, using the following techniques: N by Kjeldahl method, P and B by colorimetry, K, Ca, Mg, Na, Fe, Mn, Cu and Zn by atomic absorption, and S by turbidimetry (Combatt et al., 2014).

Preparation of dry sludge (DS) and biochar (PS)
The DS was obtained from the saltpeter PTAR located in Bogota D.C. (Colombia). DS was then dried outdoors for 30 d at room temperature (30°C) in the greenhouses of the Faculty of Agricultural Sciences of the Universidad Nacional de Colombia. Approximately 5 kg of DS were used for the production of the PS biochar. A proportion of this material was pyrolyzed in the rotary kiln of the Department of Mechatronics of the Faculty of Engineering of the Universidad Nacional de Colombia, Bogota campus. The PS was obtained under rapid pyrolysis with a temperature of 850°C for 40 min and a heating rate of 80-100°C/min. The PSs were prepared according to the methodology proposed by Jeffery et al. (2011).

The obtained PS (1.5 kg) was packed in bags and stored at room temperature at 15°C as average temperature and then was taken to the Soil biology laboratory at the Faculty of Agricultural Sciences of the Universidad Nacional de Colombia, for chemical and physical characterization. Before applying the different treatments, a ball hammer and two sieves of number 10 and 20 with openings of 2.0 and 0.85 mm, respectively, were used to reduce the particle size of the two materials. For the PS, an approximate grain size of 2 mm was used. These assessments were performed considering the International Biochar Initiative and the set of characteristics that define the quality of biochar for use in agriculture. Those parameters include particle size distribution of these materials, pH, specific area, porosity, C, and nutrient content, as well as pollutant content (heavy metals and polycyclic aromatic hydrocarbons) (IBI, 2012). The chemical analysis of both PS and DS was determined under the same parameters of the soil samples. The chemical characterization of the two materials is shown in Table 1.

We performed a laboratory characterization of the two sludges of the treatments with the following variables: bulk density, cation exchange capacity (CEC), organic carbon, organic matter, ash, carbon/nitrogen ratio, pH (acidity reaction), total nitrogen, ammoniacal nitrogen, nitric nitrogen, phosphorus (P), potassium (K), calcium (Ca), magnesium (Mg), sulfur (S), sodium (Na), iron (Fe), manganese (Mn), copper (Cu), zinc (Zn) and boron (B) (Tab. 1).

Experimental design
We carried out the test in the greenhouses of the Faculty of Agricultural Sciences of the Universidad Nacional de Colombia between February and July 2015, under average daytime temperatures of 24°C and average nighttime temperatures of 18°C. The experiment was carried out in the pots. The soil mixed with the biocarbon and DS (substrate) was placed in the pots. We planted five seeds of the tomato variety Miland at a depth of 3 cm. We previously treated the seeds with hot water to counteract possible pathogens. Once
the seeds germinated after 10 d, we transplanted those with the most vigor into the center of each pot, which were randomly arranged on the greenhouse tables. We irrigated the plants by sprinkling with a flow rate of 2 L/h, three times a week, based on the needs of the crop. We subsequently sampled the plants destructively six months after sowing when they had grown eight branches; the biomass on the soil surface was collected by cutting at the base of the stem. No chemical and organic synthesis products were applied during the test. In the absence of reference levels for the doses established for both DS and PS, we based the amounts that we applied in this research on studies by Chan et al. (2007) and Van Zwieten et al. (2009), who determined benefits of biochars produced from paper sludge and poultry feces with an average dose of 10-15 t ha⁻¹.

The experimental design was associated with an incomplete factorial in a completely randomized arrangement, since the doses were nested for each material, with DS and PS materials. The nested doses were 0.0, 2.5, 5.0, 7.5, 10.0, and 15.0 t ha⁻¹ consisting of 12 randomized treatments on the experimental units (pot with a tomato plant), using three repetitions per treatment. The treatments were labeled as follows: (T₀) control (only soil for both materials); (DS₁) 2.5 t ha⁻¹; (DS₂) 5 t ha⁻¹; (DS₃) 7.5 t ha⁻¹; (DS₄) 10 t ha⁻¹; (DS₅) 15 t ha⁻¹; (PS₁) 2.5 t ha⁻¹; (PS₂) 5 t ha⁻¹; (PS₃) 7.5 t ha⁻¹; (PS₄) 10 t ha⁻¹; and (PS₅) 15 t ha⁻¹. The response variables evaluated simultaneously were dry matter (g), and height (cm) of the plants measured six months after planting. Finally, with the same treatment structure, we evaluated the macro and microelements (N, P, K, Ca, Mg, Cl, Fe, Mn, Cu, and Zn) in leaves 6 months after sowing.

**Statistical analysis**

The relative change was analyzed taking into account the final values of the variables of PS compared to the initial values of DS, represented by a radial diagram. An exploratory analysis of the data was carried out to describe the variables involved in the study statistically. The analysis included calculations of the mean, standard deviation, and coefficient of variation of the two response variables (height and plant weight), discriminated by treatment. A conditional graph was constructed to show the relationships of all the variables involved.

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**TABLE 1. Evaluated properties of dry and pyrolyzed sludge.**

| Variables                     | Dry sludge   | Pyrolyzed sludge | Relative change (%) |
|-------------------------------|--------------|------------------|---------------------|
|                               | Units        | Dry base         | Dry base            |                    |
| Bulk density                  | g cm⁻³       | 0.88             | 0.79                | -10.23 (⁺⁻)        |
| Cation exchange capacity (CEC)| cmol kg⁻¹    | 28.17            | 64.45               | 128.79             |
| Organic carbon                | %            | 20.4             | 19.31               | -5.34              |
| Organic matter                | %            | 44.26            | 41.9                | -5.33              |
| Ashes                         | %            | 55.74            | 58.1                | 4.23               |
| Carbon/Nitrogen ratio         | w/w          | 8.5              | 8.05                | -5.29              |
| pH (acidity reaction)         |              | 5.8              | 5.9                 | 1.72               |
| Total Nitrogen                | %            | 2.4              | 2.4                 | 0.00               |
| Ammoniacal Nitrogen           | %            | 0.024            | 0.084               | 250.00             |
| Nitric Nitrogen               | %            | 0.075            | 0.123               | 64.00              |
| Phosphorus (P)                | %            | 0.93             | 1.1                 | 18.28              |
| Potassium (K)                 | %            | 0.21             | 0.35                | 66.67              |
| Calcium (Ca)                  | %            | 1.900            | 2.8                 | -99.85             |
| Magnesium (Mg)                | %            | 0.24             | 0.27                | 12.50              |
| Sulfur (S)                    | %            | 0.62             | 2.1                 | 238.71             |
| Sodium (Na)                   | %            | 0.056            | 0.062               | 10.71              |
| Iron (Fe)                     | ppm          | 4183             | 17372               | 315.30             |
| Manganese (Mn)                | ppm          | 179              | 266                 | 59.78              |
| Copper (Cu)                   | ppm          | 29.6             | 37.2                | 25.68              |
| Zinc (Zn)                     | ppm          | 360              | 777                 | 115.83             |
| Boron (B)                     | ppm          | 14               | 13.5                | -3.57              |

(⁺) Values given by 100*(end value - start value)/start value. The sign (⁻) means a decrease.
A bivariate analysis of variance was used for the evaluation of the effects of the added material and its nested doses on the height and weight of the plant, for the incomplete factorial experimental design in a completely randomized arrangement. We validated the assumptions of bivariate normality and equality of matrices of variances and covariances by treatments using the Royston test and the Box M test, respectively. The statistical package R was used for all analyses. For the analysis of the macro and micro nutritional elements of the leaf, the relative percentage change of the treatment was measured with a higher average compared to the control.

Results and discussion

Under the conditions of this study, the apparent density showed a 10.23% reduction in the conversion of DS to biochar or PS, which could be caused by the micropores that are generated when the temperature in the pyrolysis process increases (Verheijen et al., 2009). The pH level remained stable in the conversion from DS to PS, despite rapid pyrolization (high temperature and short residence times), which is thought to increase alkalinity due to the loss of carboxylic acid groups by temperature (Harris and Tsang, 1997). This situation could be due to the fact that there was no significant change in the amounts of organic carbon in this conversion, i.e., there would not be a noticeable loss of these functional groups. In addition, higher levels in the contents of P and K were found in treatments with PS as well as a greater positive percentage difference for these treatments (Tab. 1).

There could be a significant contribution of P from protein materials and polyphosphates (which were not lost in the pyrolysis process) from detergents that might be constituents of the evaluated DS (Korboulewsky et al., 2002; Esteller et al., 2009). However, the levels of total N, Mg, and Na were similar for the two materials (with increases not greater than 13% in the percentage change in the conversion from DS to PS). The permanence of the levels of these elements after pyrolysis in the PSs could be due to the fact that the temperatures did not reach the point of volatilization. In contrast, Gaskin et al. (2008) found the volatilization of N in sludge when the temperature was increased from 300 to 700°C.

In general, the percentages of micronutrients were higher in the PS, while organic matter showed higher levels in the DS. In contrast, authors such as Okuno et al. (2005) consider that Ca begins to decrease in the biochar matrix in pyrolysis with temperatures above 600°C. The higher values of Ca in the PSs in this study may be associated with higher levels in the height and dry matter of the plant. With the addition of PS, there is an increase in water retention capacity and a greater uptake of nutrients due to an increase of the CEC (Fig. 1).

Fe, Cu, and Zn increased with the conversion from DS to PS in percentages of 315.35%, 25.6%, and 115.8%, respectively (Tab. 1). Some studies corroborate the results of our research, in which authors such as Lehmann et al. (2006) establish that, although nutrients such as N and P in the sludge may decrease due to volatilization during pyrolysis, heavy metals, such as Cu and Zn, can increase their concentration. In studies by Hossain et al. (2010), different heavy metals such as Zn, Pb, and Ni increased their contents in sludge pyrolysis at temperatures greater than 500°C. This implies that the increase of these elements with PS in the substrate can affect the chemical properties of the sludge, so that there is a competition of nutrients with a high amount of Fe.

Regarding the CEC, a relative change (increase) of 128.8% was found for the PS with reference to DS. This result is contrary to the expected decrease in the values of this variable, as reported by Lin et al. (2012) and Rajkovich et al. (2012), and could happen due to the removal of functional groups that increase ion retention with their load.
The higher level of CEC in the PSs in this research could be due to a high number of these groups and also to the conservation of heteroatoms within them (Cantrell and Martin, 2012).

With increasing doses of both PS and DS in each treatment, there was a tendency to increase the dry weight and height of the plants (Fig. 2). The bivariate analysis of variance showed no significant effect attributable to the PS and DS materials on the height or the dry matter \((P=0.788)\) of the plants but rather between doses of each material \((P<0.001)\) (Tabs. 2 and 3).

**TABLE 2.** In-treatment dose test statistics.

| Statistics      | Value | F-Value | Num DF | Den DF | Pr>F |
|-----------------|-------|---------|--------|--------|------|
| Wilk’Lamda      | 0.979 | 0.24    | 2      | 23     | 0.788|

**TABLE 3.** Inter-material dose test statistics.

| Statistics      | Value | F-Value | Num DF | Den DF | Pr>F |
|-----------------|-------|---------|--------|--------|------|
| Wilk’Lamda      | 0.003 | 9.21    | 20     | 46     | <.0001|

In both DS and PS, the control was below average compared to the other treatments. Taking as reference the treatments with higher plant heights (DS\(_5\) and PS\(_5\)), the control plants were found to be 27.63% and 30.92% shorter than PS\(_5\) and DS\(_5\) plants, respectively (Tab. 4). As in the present study, Hossain et al. (2001) reported significant effects on the height of cherry tomato plants with applications of biochar made with sludge from treatment plants in the Sydney area (Australia). In that study, applications of biochar were evaluated in the greenhouse, both alone and in mixtures with inorganic fertilizers, using doses of 10 t ha\(^{-1}\). Likewise, Silva et al. (2017) found similar results in seedlings of _Eucalyptus grandis_ W, evaluating DS and PS.

In accordance with the present research, Cabrera et al. (2007) also found a positive effect on the height of tomato plants with the addition of DS to the soil from a WWTP, compared to soils fertilized with urea and soils without fertilization in the greenhouse. Such results can be attributed to improvements in some chemical properties of soils.

**TABLE 4.** Descriptive statistics for response variables height (cm) and dry weight (g) of plants.

| Variables   | Dry sludge | Pyrolyzed sludge |
|-------------|------------|------------------|
| Dose        | 0  2.5  5  7.5  10  | 0  2.5  5  7.5  |
| Height      | \(\bar{x}\) **16.100** 20.900 28.733 31.066 39.200 58.266 16.100 20.766 29.566 35.800 42.466 52.066  |
|             | **S** 2.080 4.300 3.611 3.827 6.148 3.295 2.080 3.234 2.274 6.791 6.143 5.772  |
|             | **CV** 12.924 20.574 12.569 12.321 15.686 5.656 12.924 15.576 7.692 18.969 14.466 11.087  |
| Dry weight  | **\(\bar{x}\) 6.217** 9.260 16.715 21.071 30.678 49.717 6.217 7.913 15.922 25.221 30.478 46.175  |
|             | **S** 2.883 2.065 2.692 2.316 3.397 5.767 2.883 1.012 1.060 7.519 5.719 7.839  |
|             | **CV** 46.37 22.30 16.109 10.991 11.074 11.600 46.378 12.792 6.662 29.812 18.766 16.977  |

*\(\bar{x}\): average, **S: standard deviation, **CV: coefficient of variation.*

**FIGURE 2:** Dry weight distribution as a function of plant height per nested dose in each experimental treatment.
produced by an application of these materials, due to the contribution of immediate and gradual inorganic forms of organic molecules.

In this research, dry matter and plant height increased with the addition of higher doses of both LS and LP. The control plants (with an average of 6.2 g of dry matter) were found to have 12.5% and 13.2% less dry matter than the DS, and PS, plants, respectively (Tab. 4). This could be due to the fact that these types of sludge contain high amounts of nutrients and organic matter that could have a positive effect and contribute to a greater gain of dry matter given higher doses of these materials (Melo et al., 2007). In this regard, Moral et al. (2005) also state that higher production of plants may be due to a greater amount of organic matter present in the biochar. These organic elements are composed of proteins, simple sugars, organic acids, amino acids, and peptides that are easily degraded by microorganisms, increasing those populations that contribute to a greater dynamic of transformation from non-soluble elements to elements available to plants. The addition of PS could cause greater production and growth in the plant by the reduction in the leaching of nutrients and the increase in the union of organic matter and nutrients through the cation exchange capacity (Amonette, 2009; Granatstein et al., 2009; Atkinson et al., 2010; Lehmann et al., 2011).

In other studies, in tomato, Hossain et al. (2010) found no significant differences with the addition of PS at doses of 10 t ha⁻¹ and inorganic fertilizers. However, the researchers reported higher plant growth with amendments added to sludge biochar mixtures and chemical fertilizers. In the present research, with the comparative analysis of the application of the two materials, the profiles and effects were similar to the response in dry weight (g) and height (cm) of the evaluated tomato plants (Fig. 2).

The increase in dry matter and height in the plants with the addition of PS in this study could be related to the contribution of these materials to generate better conditions and dynamics of the different properties of the substrates and, therefore, improve plant growth. In this sense, studies such as those by Glaser et al. (2002) describe improvements in the different properties of the soil with the addition of biochar or pyrolysis and the increase of dry matter in plants. The researchers argue that terra preta soils have approximately 18% higher water retention values, compared to adjacent soils. Thus, the addition of any organic amendment produces an increase in water retention, which will be represented as a positive increase in the different properties of the soil, and consequently, in improved nutrient uptake by plants. On the other hand, Hue and Ranjith (1994), McBride et al. (1997), and Shuman (1998) state that dry sludge can increase nutrient availability due to the low molecular weight of its aliphatic components and the increase of CEC in the soil.

Comparing the application of the DS and PS carried out according to the norms of this study, the use of biochar from PS could be a better option because of the stabilization of this material that occurs in its conversion to biochar through the pyrolysis process. For example, the use of PS can reduce the availability of heavy metals included in the soil, which pass into the trophic chain through the root absorption of plants (Mtshali et al., 2014). Although present at low values, the increase of metals such as Fe and Cu in the conversion of dry sludge to biochar in percentages of 3.1 and 0.2%, respectively, was evident in this research (Tab. 1).

**Foliar analysis**

According to the different percentages of nutrients from the foliar samples found in the various treatments compared to the control, generally, higher levels of micronutrients were found in the PS treatments. However, higher values were found in the macronutrient levels of the DS (except for K) (Tab. 5).

The absorption of N by tomato plants was present at higher levels within the PS treatments for treatments with higher doses of this material. For example, treatment LP₅, with the highest dose of PS in the substrate, showed the highest concentration of N in plant tissues in the pyrolyzed treatments. This could be due to the concentration of N in the ashes of the biochar adsorbed by the plant from the substrate. In this sense, authors such as Steiner et al. (2008), Zimmerman (2011), Lehmann et al. (2010) and Rajkovich et al. (2012) suggest that the uptake of nutrients such as N increases in biochar that comes from high temperatures because they contain a lower labile fraction. This means that there is less immobilization by microorganisms and, therefore, there is a greater amount of N available to be absorbed by the plant. Among the PS treatments, the PS₅ showed the highest level of Ca in the leaf tissue of the plants (Tab. 5), which could be correlated with a higher level of biochar for this treatment, due to the initial high concentration of Ca in the added PS material (Tab. 1).

However, the function of K is associated with root growth, tolerance to water stress, cellulose formation, enzymatic activity, and photosynthesis (Thomson, 2008). Although the treatments with higher doses of the two materials in this study had the highest yields (dry matter) and heights (DS₅ and PS₅), the same was not found regarding K uptake.
TABLE 5. Foliar analysis of nutritional elements in different treatments with applications of dry sludge and sludge biochar in Solanum lycopersicum L.

| Treatments | Macro elements (g*100 g⁻¹) | Micro elements (g*100 g⁻¹) |
|------------|-----------------------------|----------------------------|
|            | N  | P   | K  | Ca | Mg | S  | Na | Fe | Mn | Cu | Zn | B |
| T₀         | 0.16 | 6.51  | 0.72  | 4.86  | 1.20  | - | - | 42.8 | 1.93 | 0.31 | 0.36 | 0.51 |
| DS₁        | 2.25 | 0.19  | 3.50  | 1.90  | 0.63  | 0.52 | 0.129 | 3793.0 | 193.0 | 9.7 | 67.9 | 17.7 |
| DS₂        | 1.49 | 0.20  | 2.80  | 1.40  | 0.34  | 0.41 | 0.059 | 14036.0 | 150.0 | 5.2 | 44.7 | 15.3 |
| DS₃        | 1.11 | 0.20  | 2.40  | 1.50  | 0.25  | 0.39 | 0.055 | 3693.0 | 121.0 | 8.1 | 39.4 | 16.1 |
| DS₄        | 2.26 | 0.24  | 2.70  | 1.60  | 0.42  | 0.41 | 0.073 | 7608.0 | 184.0 | 8.9 | 71.6 | 19.1 |
| DS₅        | 1.68 | 0.22  | 3.10  | 1.90  | 0.45  | 0.48 | 0.0071 | 4448.0 | 164.0 | 8.8 | 58.4 | 20.2 |
| PS₁        | 1.40 | 0.24  | 3.30  | 1.70  | 0.55  | 0.39 | 0.150 | 4786.0 | 178.0 | 7.4 | 62.1 | 16.0 |
| PS₂        | 1.86 | 0.24  | 3.00  | 1.80  | 0.52  | 0.31 | 0.081 | 14036.0 | 150.0 | 5.2 | 44.7 | 15.3 |
| PS₃        | 1.87 | 0.22  | 3.80  | 1.40  | 0.53  | 0.35 | 0.094 | 3769.0 | 121.0 | 8.1 | 39.4 | 16.1 |
| PS₄        | 1.33 | 0.24  | 3.40  | 1.40  | 0.32  | 0.38 | 0.056 | 7608.0 | 184.0 | 8.9 | 71.6 | 19.1 |
| PS₅        | 2.03 | 0.24  | 2.80  | 1.90  | 0.45  | 0.38 | 0.077 | 4448.0 | 164.0 | 8.8 | 58.4 | 20.2 |
| Minimum optimal level | 2.80 | 0.30  | 3.50  | 1.60  | 0.36  | 0.64 | 0.05 | 84.0 | 55.0 | 6.0 | 40.0 | 54.0 |
| Maximum optimal level | 4.20 | 0.45  | 5.00  | 3.20  | 0.49  | 1.94 | 0.2 | 112.0 | 65.0 | 15.0 | 60.0 | 76.0 |

TABLE 6. Relative changes for different elements in foliar tissue samples for treatments and evaluated doses.

| Treatments | N  | P    | K  | Ca  | Mg  | Cl  | Fe  | Mn  | Cu  | Zn  |
|------------|----|------|----|-----|-----|-----|-----|-----|-----|-----|
| DS₁        | 1306.3 | -79.7 | 386.1 | -60.9 | -47.5 | 8762.1 | 9900 | 3029.0 | 18761.1 | 3370.5 |
| DS₂        | 831.3 | -96.9 | 288.9 | -71.2 | -71.7 | 32694.4 | 7672.0 | 1577.4 | 12316.7 | 3056.9 |
| DS₃        | 593.8 | -96.9 | 233.3 | -69.1 | -79.2 | 8528.5 | 6169.4 | 2512.9 | 10844.4 | 3056.9 |
| DS₄        | 1312.5 | -96.3 | 275 | -67.1 | -65 | 17675.7 | 9433.7 | 2771.0 | 19788.9 | 3645.1 |
| DS₅        | 950 | -96.6 | 330.6 | -60.9 | -178.6 | 10292.5 | 8397.4 | 2738.7 | 16122.2 | 3860.8 |
| PS₁        | 812.5 | -96.3 | 358.3 | -65.0 | -54.2 | 11082.2 | 9122.8 | 2287.1 | 17150 | 3037.3 |
| PS₂        | 1062.5 | -96.3 | 316.7 | -63.0 | -56.7 | 14234.1 | 10936.3 | 2351.6 | 17205.6 | 3311.8 |
| PS₃        | 1068.8 | -96.6 | 427.8 | -71.2 | -1.9 | 8706.1 | 8708.3 | 4029.0 | 24011.1 | 5507.8 |
| PS₄        | 731.3 | -96.3 | 372.2 | -71.2 | -73.3 | 4956.1 | 4734.2 | 2512.9 | 18094.4 | 4056.9 |
| PS₅        | 1168.8 | -96.3 | 288.9 | -60.9 | -62.5 | 8486.4 | 8708.3 | 3900 | 25872.2 | 6390.2 |

in plant tissues in these same treatments. In fact, the lowest amounts of K absorbed were found in treatments DS₅ and PS₅ (Tab. 6). This reaction can be possibly due to the long-term action of biochar, as described by Major et al. (2010) in studies conducted in corn crops between 2003 and 2006, where there were increases in elements such as K, Ca and Mg until the third year, with the addition of biochar made from wood residues. However, it is expected that the amounts of Ca and Mg contributed by biochar in the soil will increase in the long term, which may have an impact on leaf tissues through the uptake of the elements by the plant, as established by authors such as Major et al. (2010).

It is known that P is usually conserved during the volatilization of associated organic molecules and is present as ashes within the biochar, which when solubilized, becomes available to plants (Mašek and Brownsort, 2011; Escalante-Rebolledo et al., 2016). However, McLaughlin et al. (2009) demonstrated that P in the sludge acts as a slow-release fertilizer in soils deficient in this element (Hossain et al., 2010). In addition, the concentration of P in the sludge can be reduced by the conditions of pyrolysis through the evolution of volatile compounds of greater molecular weight, while the concentration of metals present in the biochar is expected to increase (Hossain et al., 2010).

This could explain the results of this study regarding the lower content of P in both DS and PS compared to the control and the higher Fe content found in these two materials. In this research, although the amounts of P are significantly lower in the two sludge materials compared to the control (Tab. 5), it could be inferred that in T₀, the element is neutralized (due to high amounts of bases like K). For this reason, it is not available for the plant and could not provide a contribution to its production (interpreted in terms of height and dry matter).
Na ions in large quantities can affect the seeds of S. lycopersicum by decreasing the water potential in the substrate (El-Habbasha et al., 1996; Cuartero and Fernández-Muñoz, 1998). These concentrations also affect the growth of the roots of the plant by altering the absorption of water (an osmotic component) (Shannon and Grieve, 1999). The above could be reflected in this research given the results obtained regarding the high content of Na since higher levels of this element were found in the leaf samples of the treatments with lower height and dry matter (DS, and PS), and lower levels of Na in treatments with higher height (DS5 and PS5) (Tab. 4).

In the analysis of Mn, Zn, and B, there was no clear trend that could relate a higher dose of treatments to an increase in the content of these elements in plant tissues. Foliar analyses revealed significant amounts of Fe, particularly in DS in treatments DS2 and DS4. The PS treatments that were found with greater amounts of this element were the PS1 and PS2. In relation to this, Pérez-Sanz et al. (2002) investigated the efficiency in the growth of citrus crops with the application of sludge enriched with Fe without finding significant improvements in this growth parameter. These results are in line with what was obtained in our study, particularly with the application of treatments that include biochar, since it was found that treatments with high amounts of Fe such as PS1 and PS2 had the lowest plant height and dry matter values.

Conclusions

The results obtained indicated that the applications of DS and PS on sludges with tomato plants under greenhouse conditions generated an effect on the height and dry matter of the plants, which was proportional to the increase in the doses of these materials to the substrate. Although no significant differences were found in the application of the two materials, there were statistical differences found in the nested doses for each material, and greater response in the evaluated variables at higher doses of each material.

With the addition of PS to the substrate, an improvement in the availability of nutritional elements of the soil, such as K was observed. Additionally, an increase in the absorption of these elements by tomato plants (particularly in the PS5) was observed, which positively influenced greater heights and values of dry matter obtained in the plants. Essential elements such as Ca increased in the substrate with the application of DS and PS. However, increased amounts of Fe were also found in the PS, which could have a negative impact on the chemical conditions of the soil and plants.

When higher levels of nutrients (particularly nitrogen) are found in the plants under treatment with PS, a contribution of nutritional elements by this material was inferred. This is a very important aspect that must be investigated in depth due to its potential benefits in the nutrition and production of cash crops.

The application of DS and PS could be a viable proposal for agricultural systems as an alternative to the disposal of sewage sludge and as an amendment or soil conditioner. Although greater benefits were found in some variables measured in this study with the application of some DS, there could be a greater benefit in environmental terms with the use of PSs given that the heavy metal load is reduced when pyrolysis is applied to the soil. This benefit may mean a lower impact in terms of bioaccumulation of these metals in the plant.

However, for future studies, it is necessary to analyze the content of toxic organic compounds such as dioxins, furans, phthalic acid esters, polycyclic aromatic hydrocarbons, and other organic contaminants that may be present in DS and PSs. This information can be important to define criteria for the selection of one of these materials to be used as a soil improver.

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