LETTER

Satellites may underestimate rice residue and associated burning emissions in Vietnam

Kristofer Lasko1,5, Krishna P Vadrevu2, Vinh T Tran3, Evan Ellicott1, Thanh T N Nguyen4, Hung Q Bui3 and Christopher Justice1

1 Department of Geographical Sciences, University of Maryland, College Park, MD 20742, United States of America
2 Earth Science Office, NASA Marshall Space Flight Center, Huntsville, AL, United States of America
3 Faculty of Information Technology, Hanoi Pedagogical University 2, Vinh Phuc, Viet Nam
4 University of Engineering and Technology, Vietnam National University Ha Noi, Ha Noi, Viet Nam
5 Author to whom any correspondence should be addressed. E-mail: klasko@umd.edu

Keywords: Rice straw, PM2.5, GFED, REAS, biomass burning, southeast Asia, bottom-up

Abstract

In this study, we estimate rice residue, associated burning emissions, and compare results with existing emissions inventories employing a bottom-up approach. We first estimated field-level post-harvest rice residues, including separate fuel-loading factors for rice straw and rice stubble. Results suggested fuel-loading factors of 0.27 kg m\(^{-2}\) (±0.033), 0.61 kg m\(^{-2}\) (±0.076), and 0.88 kg m\(^{-2}\) (±0.083) for rice straw, stubble, and total post-harvest biomass, respectively. Using these factors, we quantified potential emissions from rice residue burning and compared our estimates with other studies. Our results suggest total rice residue burning emissions as 2.24 Gg PM2.5, 36.54 Gg CO and 567.79 Gg CO\(_2\) for Hanoi Province, which are significantly higher than earlier studies. We attribute our higher emission estimates to improved fuel-loading factors; moreover, we infer that some earlier studies relying on residue-to-product ratios could be underestimating rice residue emissions by more than a factor of 2.3 for Hanoi, Vietnam. Using the rice planted area data from the Vietnamese government, and combining our fuel-loading factors, we also estimated rice residue PM2.5 emissions for the entirety of Vietnam and compared these estimates with an existing all-sources emissions inventory, and the Global Fire Emissions Database (GFED). Results suggest 75.98 Gg of PM2.5 released from rice residue burning accounting for 12.8% of total emissions for Vietnam. The GFED database suggests 42.56 Gg PM2.5 from biomass burning with 5.62 Gg attributed to agricultural waste burning indicating satellite-based methods may be significantly underestimating emissions. Our results not only provide improved residue and emission estimates, but also highlight the need for emissions mitigation from rice residue burning.

Introduction

Crop residue burning is an important source of greenhouse gases and aerosols (Streets et al 2003, Crutzen and Andreae 1990). The burning of crop residues contributes to at least 34% of global biomass burning emissions (Streets et al 2004). While these and other studies provide useful general estimates, analyses need to be region-specific to enable emissions mitigation. Of the different crop residues, rice residues are prevalently burned in South/Southeast Asian countries in addition to forest biomass burning (Streets et al 2003, Biswas et al 2015).

Rice (Oryza sativa) is the staple crop for livelihood in Southeast Asia and more specifically, Vietnam. During 2015, Vietnam produced 45.2 million metric tonnes of rice with most production in the Mekong River Delta (57%) and the Red River Delta (15%) (Vietnam Government 2015). The Red River Delta is home to Hanoi, the capital of Vietnam which outside of the immediate downtown area, exhibits a mosaic landscape dominated by paddy rice, small-holder...
farms, and plantations, all intermixed amongst a growing peri-urban area (Pham et al. 2015). Thus, in Hanoi, many residential and commercial areas are not only impacted by urban emissions, but also by smoke from rice residue burning. Studies have attributed crop residue burning to local and regional impacts including long-range transport with effects persisting for weeks or months impacting air quality, atmospheric chemistry, weather, and biogeochemical cycles (Badarinath et al. 2007, 2009, Vadrevu et al. 2012, 2014, 2015, Cristofanelli et al. 2014, Reddington et al. 2014, Ponette-González et al. 2016, Yan et al. 2006, Le et al. 2014). For Hanoi in particular, nocturnal radiation inversions occur during the October rice harvest and burning, greatly enhancing the negative air quality impact of fine-particulate matter emissions (Hien et al. 2002).

Around Hanoi, the typical paddy rice field size ranges from 150–2280 m² (Patanothai 1996) with an average of 790 m² (σ = 625 m²) (this study). Hanoi is located within the heart of the Red River Delta which is Vietnam’s second largest rice producing hub with over 35% of the land dedicated to rice (Vietnam Government 2015). Rice is routinely irrigated and double-cropped in Hanoi Province with winter rice harvested during June and spring rice harvested during late September–October. After harvest, a large volume of rice straw is left in rows or piles on the field as well as uncut stubble (figures 1(a) and (b)). In order to prepare the field for the next harvest, farmers routinely burn the residues. Typical burning of a rice straw pile and post-burned field in Hanoi are shown (figures 2(a) and (b)). Mostly, the straw is burned or reincorporated into the soil while some is used for cattle feed, cook stoves, composting, and mushroom cultivation (Trach 1998, Hong Van et al. 2014, Duong and Yoshiro 2015, Nguyen 2012, Oritate et al. 2015). In comparison to the rural areas, suburban areas such as Hanoi typically burn a higher proportion of rice straw as these areas have fewer cattle relying on it for food (Duong and Yoshiro 2015). Thus, with a higher proportion of residue burned in Hanoi, there is amplified impact from emissions. Further, post-harvest rice straw is assumed to have moisture content of about 15% or less, however this varies depending on conditions, and residue structure/density can also have an impact on resulting emissions (Korenaga et al. 2001, Arai et al. 2015). In order to estimate emissions impact, accurate bottom-up quantification of residue production and burning is needed.

Earlier studies on estimating emissions from crop residue burning have used agricultural production data, a crop specific residue-to-product ratio, an estimate of the proportion of residue subject to burning, emission factors, and a combustion factor (Streets et al. 2003, Yevich and Logan 2003, Yan et al. 2006,

![Image](image1.png)

**Figure 1.** (a) Typical machine-harvested field in Hanoi province with dry rice straw laid in neat rows; (b) typical rice straw pile prior to burning near Hanoi City.

![Image](image2.png)

**Figure 2.** (a) Rice straw pile burning. Piles are often burned nearby to the harvested field. (b) Typical post-burned rice field in Hanoi Province, Vietnam. Most straw is burned efficiently, however much stubble is left incompletely combusted.
Cao et al. 2008, Gadde et al. 2009, Kanabkaew and Oanh 2011, Vadrevu and Lasko 2015, Zhang et al. 2015). While these studies yield insight on emissions estimation, they can be improved by incorporating field-based locally/regionally estimated fuel-loads or emissions factors established from field measurements (Oanh et al. 2011, Kanokkanjana et al. 2011, Rajput et al. 2014, Hong Van et al. 2014, Arai et al. 2015). We note that comprehensive province-level field-estimates of rice straw and rice stubble have yet to be generated for northern Vietnam.

Field studies estimating rice straw, stubble, and total post-harvest biomass production are labor intensive and costly. Accordingly, remote sensing with its synoptic and consistent coverage can be used for estimating these factors. In Southeast Asia and Vietnam, forecasting of rice yield or biomass has been done using X-band synthetic aperture radar (SAR) (Gebhardt et al. 2012, Bouvet et al. 2014), C-band SAR (Ribbes and Le Toan 1999, Chakraborty et al. 2005, Lam-Dao et al. 2009, Inoue et al. 2014, Karila et al. 2014), and L-band SAR (Zhang et al. 2009, Torbick et al. 2011). Further details on these and related mapping applications are available in recent reviews (Kuenzer and Knauer 2013, Mosleh et al. 2015, Dong and Xiao 2016). Developing a relationship between field-estimated rice biomass and SAR signal is useful for upscaling field studies to broader regions and time periods. Thus, it contributes to systematic and operational monitoring of rice residue production useful for not only estimating emissions from burning, but also emissions mitigation such as bioenergy generation.

In this study, we develop and assess a straightforward method for efficient and accurate field estimation of rice residue fuel-loading factors for straw, stubble, and total post-harvest biomass. We use these as inputs to calculate resulting potential residue burning emissions for Hanoi, Vietnam and compare results using fuel-loading factors from other studies. Using our fuel-loading factors and those from other studies we upscale our results to the entirety of Vietnam and compare with an existing emissions inventory to assess rice residue burning contribution to total emissions from all sources. We also compare our emission estimates with the satellite-derived estimates from the Global Fire Emissions Database (GFED). We then explore the potential forecasting of rice residue using field and SAR data.

We specifically address the following questions: (1) how much rice straw, stubble, and total post-harvest biomass is left in the field, and how does this compare to other regional studies? (2) What are the resulting rice residue burning emissions for major pollutants (PM$_{2.5}$, CO, and CO$_2$) in Hanoi Province? (3) What are the resulting rice residue burning emissions for the entire Vietnam and how do they compare to the existing emissions inventories from different sources? (4) How well does SAR data enable forecasting of post-harvest biomass? We addressed these questions using a field and remote sensing based approach representative of a typical double-cropped rice region in Hanoi Province, Vietnam.

Datasets and methods

SAR data

Sentinel-1 carries a 12 m long C-Band SAR yielding a unique ability to penetrate most cloud coverage. Sentinel-1 is a constellation of two satellites including Sentinel-1A with data since October 2014 and Sentinel-1B launched in spring of 2016. The constellation has a repeat-pass of approximately six days for many regions. From the Alaska Satellite Facility, we obtained a single ground-range detected, interferometric wide-swath (IW), dual-polarized image from Sentinel-1A just prior to harvest for Hanoi Province on September 18, 2016. We processed the imagery following detailed guidelines from European Space Agency using the freely-available S1 toolbox (Zühlke et al. 2015) including thermal noise reduction, multi-looking azimuthal compressions to 20 m spatial resolution to reduce noise, terrain correction using the SRTM 30 m DEM, speckle filter, and radiometric adjustments to correct for viewing geometry effects (Liang 2005). We first generated a rice map from a full time-series of Sentinel-1A time-series imagery for 2016 using a support vector machine classification method (figures 3(a) and (b)) and validated it using field data and fine-resolution imagery from Google Earth and field photos; the resulting map had an overall accuracy of 94.3% and approximately 220 000 ha of rice land area. The resulting rice map built upon results from related studies (Torbick et al. 2017, Nguyen et al. 2015) and was used to delineate rice from non-rice for field sampling (figure 3(b)).

The field data for generating the fuel-loading factors were collected from Hanoi Province and include georeferenced photos of rice and non-rice areas used in our rice area mapping classification training or accuracy assessment. For relating our field-level fuel-loading data to SAR, we used the Sentinel-1 VH-polarized image obtained approximately 2 weeks prior to harvest (September 18, 2016). The field data collection is described in the subsequent section.

Rice residue and emissions estimation

We developed a field-based approach to estimate total post-harvest rice residues including straw and stubble left in the field for burning. Using two-stage cluster sampling with the validated rice map generated from Sentinel-1A time-series imagery (figures 3(a) and (b)) we divided the study area into 5 km equal-area grid cells and randomly selected 13 rice-containing cells where four randomly selected machine-harvested fields were sampled in each cell for a total of 52 field
samples (C.I. = 9.6%, C.L. = 95%) (Sutherland 2006) (figure 3(b)). Within each harvested field, we measured the total straw and stubble weight in four randomly selected 0.5 m × 0.5 m quadrats using a Salter-Brecknell digital scale, bag with known weight, and an Extech MO290 moisture device for relative moisture content. Stubble was cut at the base before weighing, and the weight of the bag was subtracted from the measurements. We also collected ancillary data including field length, width, and number of straw rows. We estimated the amount of straw per square meter for a given field based on the following (equation (1)):

$$Q_{rs} = \frac{(RS_w \times (1 - RS_m)) \times SR \times R_l}{A}$$

where $Q_{rs}$ is the quantity of dry rice straw in kg m$^{-2}$ for a given field, $RS_w$ is the wet rice straw weight per linear meter of a straw row averaged from the quadrat measurements, $RS_m$ is the average field-measured relative moisture content, $SR$ is the number of straw rows in the field, $R_l$ is the length of the straw rows in meters, and $A$ is the field area in m$^2$.

We estimated the amount of rice stubble for a given field based on a similar (equation (2)):

$$Q_{sr} = SR_w \times (1 - SR_m) \times A$$

where $Q_{sr}$ is the quantity of dry rice stubble in kg for a given field, $SR_w$ is the average weight of rice stubble per m$^2$ from the quadrat measurements, $SR_m$ is the average measured rice stubble relative moisture content, and $A$ is the area of the field measured in m$^2$. The resulting $Q_{sr}$ yields a maximum rice stubble fuel-loading factor in kg. We also measured the number of plants in each quadrat for a measure of rice planting density.

Thus, we have three separate fuel-loading factors: straw, stubble, and total post-harvest biomass factors. Using these fuel-loading factors we estimated rice residue burning emissions based on three scenarios (1) all rice straw is burned; (2) all rice straw and stubble are burned; (3) most-likely amount burned based on the previously-mentioned surveys from the literature and our field experience. We compared these estimates to the typical approach used in most studies which relies on a residue-to-product ratio from government data on crop production. In addition, we also compared maximum potential burning estimates derived using fuel-loading factors from two smaller-scale field studies in villages in Thailand, and factors from another small-scale study in Can Tho, Vietnam located in the Mekong Delta.

We calculated the maximum potential emissions from rice residue burning based on the following (equation (3)):

$$E_a = A \times FL \times EF_a \times PB \times CF$$

where $E_a$ is the maximum potential rice residue burning emissions for a given pollutant in gigagrams, $A$ is the paddy rice planted area in hectares based on the Vietnam government statistics, $FL$ is the fuel-loading factor or that estimated from crop production data in kg ha$^{-1}$, $EF_a$ is the emissions factor for a given pollutant species in g kg$^{-1}$, $PB$ is the proportion of residue subjected to burning (from 0%–100%, i.e. residue left in the field to be burned), and $CF$ is the combustion factor indicating the burn completeness (from 0%–100%) for the residues subjected to
burning. For all scenarios estimating emissions in Hanoi Province, we selected a combustion factor of 0.8 (Aalde et al 2006) as the best available factor representative of croplands in general, the rice area in Hanoi Province based on Vietnam government statistics of 200.8 ha (Vietnam Government 2015), the emission factors for each species: CO 102 g kg$^{-1}$ (±33 g kg$^{-1}$), CO$_2$ 1585 g kg$^{-1}$ (±100 g kg$^{-1}$), and PM$_{2.5}$ 6.26 g kg$^{-1}$ (±2.36 g kg$^{-1}$) (Akagi et al 2011), the proportion burned as 100% for the maximum emissions scenario, and 50% straw and 10% stubble for the most-likely scenario based on the literature and our field experience (Tran et al 2014, Nguyen 2012, Duong and Yoshir o 2015). To estimate the fuel-loading factor for the production-statistics based scenario, we used a regionally-estimated residue-to-product ratio for rice straw of 0.75 (Gadde et al 2009b) combined with the production data from the government statistics (Vietnam Government 2015) resulting in a fuel-loading factor of 3803 kg ha$^{-1}$. For the other smaller-scale studies, we used their fuel-loading factors of 3600 kg ha$^{-1}$ (Kanokkanjana et al 2011), 5800 kg ha$^{-1}$ (Oanh et al 2011), and 3470 kg ha$^{-1}$ (Hong Van et al 2014). The latter factor was derived by averaging the reported seasonal fuel-loading factors of 3500 kg ha$^{-1}$ (spring–summer), 4300 kg ha$^{-1}$ (winter–spring), and 2600 kg ha$^{-1}$ (summer–autumn). For each scenario, all factors were held constant except for the fuel-loading factor and proportion burned.

As our study measurements are representative of a typical double-cropped paddy rice landscape and results from (Hong Van et al 2014) cover triple cropping in Mekong Delta Vietnam, we up-scaled the results from both studies to estimate contribution of PM$_{2.5}$ emissions from rice residue burning for the entire Vietnam. The rice planted area for 2008 was obtained from Vietnam Statistics Office (Vietnam 2015). We applied the triple cropping factors to the Mekong Delta rice area, and our study’s fuel-loading factors to the rest of the rice area in Vietnam. We compared the resulting rice burning emissions estimates with an existing emissions inventory (regional emission inventory in Asia (REAS)) for the latest available year, 2008 (Kurokawa et al 2013). We calculated Vietnam’s total emissions (equation (3)) for different scenarios including the maximum potential emissions (all residue burned), and the most-likely scenarios of straw and stubble burning based on the literature and our field experience suggesting approximately 50% of straw burned and 10% of stubble burned (Nguyen 2012, Hong Van et al 2014, Tran et al 2014, Duong and Yoshir o 2015, Oritate et al 2015).

While comparison of rice residue burning emissions with REAS yields insight into the relative contribution to total PM$_{2.5}$ emissions, this does not address how our field-derived estimates compare with available satellite-derived estimates. The latest available GFED version 4.1s at 0.25 × 0.25 degree was used to derive total biomass burning emissions for Vietnam, GFED also includes the portion contributed from agricultural waste burning (van der Werf et al 2010) based on MODIS burned area (Giglio et al 2013), and supplemented by small fire burned area (Randerson et al 2012). We calculated the emissions for the same year as REAS (2008) using the monthly datasets, and the same emission factors as our field study to maintain consistency. Accordingly, the emissions were derived using the amount of dry matter burned (DM) from GFED and multiplied with the aforementioned emission factors and summed for entire Vietnam.

We also evaluated the accuracy of our field-based straw measurements in eight additional randomly selected fields within Hanoi Province. First, following the same procedure as the initial field data collection we calculated the expected dry straw weight. To calculate the observed dry straw weight, we collected all rice straw within the field and weighed it. After factoring in the weight of the measuring bags and the average moisture content we arrived at an observed dry straw weight. We then compared this to our expected dry rice straw weight (kg m$^{-2}$) estimated from our field calculation (equation (1)). Using these observed and expected rice straw values we calculate an area-weighted root mean square error (RMSE). We also generated a residual plot to check for systematic errors. As the stubble measurement is more straightforward, we assumed the same accuracy as the straw measurement.

In addition to fuel-loading factor error, we also calculated the resulting emissions error rates through the simplified error propagation equations using reported error rates for emission factors, and fuel-loading factors while also accounting for constants with unknown error such as rice area (Harvard 2013).

Results

Fuel-loading factors

Results on field-level distribution of values for rice straw, stubble, total post-harvest biomass, straw moisture content, stubble moisture content, and field area are shown in figure (4). The average fuel-loading factors as measured were: rice straw 0.27 kg m$^{-2}$, rice stubble 0.61 kg m$^{-2}$, and total post-harvest residue 0.88 kg m$^{-2}$; table (1) lists more details including standard deviation and moisture content. The range of standard deviations suggests some spatial variability for the fuel-loading factors and higher ranges for the moisture content attributed to recent rain events. We also found an average field size of 790 m$^2$ (σ = 625 m$^2$), and average number of rice plants of 35.1 per m$^2$ (σ = 5.2). Based on the Vietnam government’s rice planted area data for 2015 and our fuel-loading factor, we estimated total straw production for Hanoi
we arrived at our content device (3% lab-estimated) and scale (0.4%) also factoring in error propagation from the moisture dry rice straw is estimated to be 0.27 kg m$^{-2}$ and estimated rice stubble (±0.033), rice stubble (±0.076), total residue (±0.083) kg m$^{-2}$.

Average Dry Fuel Load 0.27 kg m$^{-2}$, rice stubble (±0.076), total residue (±0.083) kg m$^{-2}$, moisture content (±3%).

As compared with the rice straw amount derived from the crop production statistics, our rice straw factor is slightly lower (3803 kg ha$^{-1}$ vs. 2700 kg ha$^{-1}$). However, when incorporating stubble, we account for a total biomass estimate which is not calculated in the production statistics method. We note other studies outside Vietnam found higher fuel-loading factors including 3600 kg ha$^{-1}$ and 5800 kg ha$^{-1}$ in two separate studies in Thailand (Kanokkanjana et al 2011, Oanh et al 2011), while the other study in the Mekong Delta, Vietnam had a lower average of 3470 kg ha$^{-1}$ (Hong Van et al 2014). This wide variation suggests region specific fuel-loading factors are an important base for emissions estimates.

**Emissions estimates**

In Hanoi Province, we used our resulting fuel-loading estimates for quantifying rice residue emissions. We used the cropped area data, combustion factor, and emission factors for CO, CO$_2$, and PM$_{2.5}$ as described above. We present the emissions for each scenario including: maximum emissions from typical production-statistics, maximum emissions using fuel-load factors from field-based study in Can Tho, Vietnam (Hong Van et al 2014), and maximum potential and most-likely emissions using our fuel-load factors of rice straw, and total post-harvest biomass. We also estimated maximum potential emissions using fuel-load factors from two village-level studies in Thailand (Kanokkanjana et al 2011, Oanh et al 2011) using their respective rice straw fuel-loading factors of 3600 kg ha$^{-1}$ and 5800 kg ha$^{-1}$. Our results suggest maximum emissions ranges for CO (44.24–144.19 Gg), CO$_2$ (687–2240 Gg), and PM$_{2.5}$ (2.72–8.85 Gg) in table (2).

Based on averaging the residue burning survey data
and from our field experience, the most likely proportion burned is 50% straw and 10% stubble. We find the most-likely combined straw and stubble emissions for Hanoi as CO (36.54 Gg), CO₂ (567.79 Gg), and PM₂.₅ (2.24 Gg). The individual scenario results are presented in table 2 including the error rates for each value in parentheses.

We highlight notable variation between the different scenarios. While all rice stubble is not necessarily burned in the study area, in other regions such as India both the straw and stubble are routinely burned (Gadde et al 2009, Gupta et al 2001, Vadrevu et al 2015). However, many studies estimating emissions from rice residue burning rely on residue-to-product ratios. Without accounting for rice stubble that is actually burnt on field, these and other studies may be underestimating emissions by a factor of about 2.3 (table 2).

We also estimated total rice residue burning emissions for the entire Vietnam during 2008 and compared with the existing emission inventory (Kurokawa et al 2013). Based on the most-likely
Table 2. Potential emissions for Hanoi Province, Vietnam using fuel-loading factors from the literature and our study. Gg = gigagrams. Maximum potential emissions assume 100% burned for first five listed scenarios, while the bottom three scenarios are the most-likely emissions based on survey and field data on residue proportion subject to burning.

| Scenario                                      | Fuel load (kg ha⁻¹) | PM₂.₅ Emissions (Gg) | CO Emissions (Gg) | CO₂ Emissions (Gg) |
|-----------------------------------------------|---------------------|----------------------|-------------------|--------------------|
| Crop Production Stats-based FL                | 3803                | 3.824                | 62.31             | 968.30             |
| Can Tho, Vietnam straw FL (Hong & al 2014)   | 3470 (±830)         | 3.489 (±7.759)       | 56.86 (±113.93)   | 883.51 (±10884)    |
| Can Tho, Vietnam FL (Straw & Stubble)        | 7330 (±1730)        | 7.371 (±16.327)      | 120.10 (±239.53)  | 1866.32 (±22706)   |
| This study (straw only)                       | 2700 (±330)         | 2.715 (±5.359)       | 44.24 (±76.20)    | 687.46 (±4709)     |
| This study (straw and stubble)                | 8800 (±830)         | 8.849 (±17.126)      | 144.19 (±241.99)  | 2240.61 (±12662)   |
| This study, most-likely emissions (straw)     | 2700 (±330)         | 2.715 (±5.359)       | 44.24 (±76.20)    | 687.46 (±4709)     |
| This study, most-likely emissions (stubble)   | 6100 (±760)         | 6.163 (±12.19)       | 100.00 (±17.38)   | 155.31 (±1126)     |
| This study, most-likely emissions (straw and stubble) | — 2.245 (±5.309) | 36.54 (±75.02)       | 567.79 (±3925)    |

emissions scenario (10% stubble, 50% straw burned), 75.98 Gg of PM₂.₅ are emitted from rice residue burning in Vietnam with stubble accounting for 18.36 Gg and straw 57.62 Gg. The total rice residue burning accounts for 12.8% of Vietnam's total PM₂.₅ emissions and is the 2nd highest PM₂.₅ combustion source after fuelwood burning. However, if all rice residue is burned, rice emissions could result in up to 36.49% of total PM₂.₅ emissions. We also add our emissions estimates to the original total from the emissions inventory of 519.81 Gg (Kurokawa et al. 2013) and arrive at new maximum of 818.50 Gg and most-likely total PM₂.₅ emissions of 595.79 Gg for the entire Vietnam. Table (3) contains the factors used for estimation, and the individual scenario results including error estimates for residue burning emissions.

GFED-derived PM₂.₅ total biomass burning emissions suggested 42.56 Gg ±16.0 Gg) with 5.62 Gg ±2.1 Gg) attributed to agricultural waste burning for the entire Vietnam. Accordingly, when agricultural waste burning emissions are compared with our field-derived residue burning emission estimates this suggests GFED underestimation by a factor of 13.5. Even if all biomass burning emissions (42.56 Gg) were attributed to residue burning, this would still be less than our field-derived estimates by a factor of 1.8. We also highlight the region-level amount of biomass burning, amount of crop residue burning, and associated emissions in table (4) suggesting the highest fraction of crop waste burning to occur in the Mekong River Delta and Red River Delta where rice is the predominant land cover. The spatial variation of biomass burning and emission estimates is highlighted in figure (7).

SAR data and biomass relationship

To evaluate the relationship of SAR backscatter with rice straw and total post-harvest biomass, we selected a single Sentinel-1A image just prior to harvest for the Hanoi Province. We found a moderately-weak relationship between the SAR backscatter and field-measured rice straw \((r^2 = 0.323, p < 0.01)\), while a moderate relationship was observed with the total rice biomass \((r^2 = 0.560, p < 0.01)\) (figures 8(a) and (b)). We observed a negative, linear relationship suggesting fields with lower backscatter values prior to harvest have more post-harvest biomass. The relationship is promising, however it needs more refinement in order to be useful for estimating rice biomass and rice straw production prior to harvest. Other studies have found good results using SAR to estimate different rice field properties, biomass, or yield (Lam-Dao et al. 2009, McNairn and Brisco 2004, Ribbes and Le Toan 1999, Wiseman et al. 2014, Paloscia et al. 1999, Bouvet et al. 2014, Erten et al. 2016, Inoue et al. 2014, Satalino et al. 2015).

Discussion

In this study, we focused on measuring residue amounts from machine-harvested fields. With the increasing urbanization, infrastructure, wealth, and interconnectedness, we anticipate most fields to switch from hand-harvesting to machine-harvesting (Nguyen et al. 2016). Hand-harvested fields are typically cut slightly higher from the ground, thus they have more stubble and less straw to potentially burn. Thus as machine-harvested fields become increasingly prevalent, residue burning emissions will be exacerbated, as it is more difficult to collect straw after machine harvest (Nguyen et al. 2016). Some studies ignore the stubble that is left on the ground, thus under-estimating total residue and resulting emissions. However, as undertaken in this study we quantify fuel-loading factors for straw and stubble which provides an improved assessment of rice residue burning emissions, as other studies often ignore the stubble factor. Our estimates incorporating fuel-loading factors for rice straw, rice stubble, and total rice biomass were useful in refining emissions in the Hanoi Province and the entirety of Vietnam. The results can be extended to similar small-holder rice production lands. We note that farmers may burn rice straw either in a pile or in an open-manner including the stubble and straw. Thus, in addition to variable fuel-loads from different harvest methods, the
Table 3. Potential emissions from rice residue burning for Vietnam. Comparison with the REAS emission inventory (Kurokawa et al. 2013) and contribution of rice residues to total PM$_{2.5}$ emissions are also shown. Error estimates for emissions are in parentheses. MRD = Mekong River Delta.

| Scenario                              | Rice area Mekong Delta (ha) | Rice area rest of Vietnam (ha) | Fuel load rest of Vietnam (kg ha$^{-1}$) | Fuel load Mekong Delta (kg ha$^{-1}$) | Proportion burned | PM$_{1.5}$ EF g kg$^{-1}$ | Rice burned MRD (Gg) | Rice burned rest of Vietnam (Gg) | Rice residue PM$_{2.5}$ (Gg) | Emissions inventory total PM$_{2.5}$ (Gg) | New total emissions inventory PM$_{2.5}$ (Gg) | Percent of total emissions from rice residue |
|---------------------------------------|-----------------------------|-------------------------------|------------------------------------------|----------------------------------------|-------------------|----------------------|-------------------|-------------------------------|-----------------------------|------------------------------------------|------------------------------------------|------------------------------------------|
| Entirety of Vietnam: Maximum potential emissions (100% straw and stubble burned) | 3858900                     | 3563300                       | 8800                                     | 7330                                    | 0.8               | 1                    | 6.26              | 22628.59                      | 25086.72                    | 298.70 (±87.72)                        | 519.81                                    | 818.50                                  | 36.49                                   |
| Entirety of Vietnam: Most likely Scenario (50% straw) | 3858900                     | 3563300                       | 2700                                     | 3470                                    | 0.5               | 6.26                | 3556.15            | 3848.36                      | 57.62 (±17.76)                   | 519.81                                    | 577.43                                  | 9.98                                    |
| Entirety of Vietnam: Most likely Scenario (10% stubble) | 3858900                     | 3563300                       | 6100                                     | 3867                                    | 0.1               | 6.26                | 1193.79            | 1738.89                      | 18.36 (±5.22)                   | 519.81                                    | 538.17                                  | 3.41                                    |
| Entirety of Vietnam: Most likely Scenario (above two combined) | –                           | –                             | –                                        | –                                        | –                 | –                   | –                 | –                            | –                                        | –                                        | –                                        | –                                        |

MRD = Mekong River Delta.
The orientation of the rice straw and piling might play a role in the combustion, moisture content, emissions factor, and resulting emissions impact; some of which have been explored, but further quantification is needed (Arai et al. 2015, Heinsch et al. 2016). Understanding the variation in residue management practices and how they impact emissions, may provide useful additional information on emissions mitigation.

Table 4. GFED-derived region-wise total biomass burned, total biomass burning PM$_{2.5}$ emissions, crop waste amount burned (derived from GFED subset), crop waste burning PM$_{2.5}$ emissions, as well as the percentage of crop waste burning out of total biomass burning.

| Region                  | Total biomass burned (Gg) | Total biomass burning PM$_{2.5}$ (Gg) | Crop waste burned (Gg) | Crop waste burning PM$_{2.5}$ (Gg) | Contribution from crop waste burning |
|-------------------------|---------------------------|---------------------------------------|------------------------|------------------------------------|-------------------------------------|
| Central Highlands       | 3215.18                   | 20.13                                 | 258.16                 | 1.62                               | 8.0%                                |
| Mekong River Delta      | 384.43                    | 2.41                                  | 277.01                 | 1.73                               | 72.1%                               |
| North Central Coast     | 438.20                    | 2.74                                  | 41.54                  | 0.26                               | 9.5%                                |
| North East              | 423.84                    | 2.65                                  | 26.60                  | 0.17                               | 6.3%                                |
| North West              | 925.75                    | 5.80                                  | 86.21                  | 0.54                               | 9.3%                                |
| Red River Delta         | 29.52                     | 0.18                                  | 20.40                  | 0.13                               | 69.1%                               |
| South Central Coast     | 595.82                    | 3.73                                  | 91.87                  | 0.57                               | 15.4%                               |
| South East              | 785.76                    | 4.92                                  | 96.54                  | 0.60                               | 12.3%                               |
| Total                   | 6798.50                   | 42.56                                 | 898.33                 | 5.62                               | 13.2%                               |

Figure 7. GFED-based spatial variation with region boundaries overlaid for: total biomass burned, crop waste amount burned, and resulting PM$_{2.5}$ emissions from the same.
Although the total estimated rice burning emissions may be relatively small as compared to other combustion and non-combustion sources (Kurokawa et al. 2013), rice residue burning in Hanoi is practiced two times per year, amplifying and temporally-concentrating the impact from burning. Accordingly, the air quality in Hanoi is frequently degraded as seen by local haze, and a high air quality index (Nguyen et al. 2015). Thus, emissions mitigation from rice residue burning is one critical aspect for improving air quality and human health.

Uncertainty with regard to different aspects of fuel-loading and emissions are important to characterize. Further, due to limited availability of data we highlight some future needs which could improve upon the rice stubble and straw burning emissions estimates. Accordingly, improvements could be made by: using separate combustion factors and emission factors for straw and stubble, deriving improved emissions factors specific to rice in both the Red River Delta and Mekong River Delta, improved mapped rice area estimates, region/crop specific combustion factors for straw and stubble, and machine versus hand-harvested fuel-loading factor comparisons. We note the government estimates of rice land area for Hanoi are different: government statistics (201 000 ha) versus mapped area (220 000 ha). Accordingly, these variations could have significant impact on resulting emissions estimates, especially once aggregating results to the national scale and accounting for both seasons.

Notably low GFED emission estimates could be attributed to a variety of factors such as prevalent cloud cover impacting satellite observations (Wilson and Walter 2016), small size of agricultural fires in Vietnam, ephemeral agricultural fires, active field management, or burning after MODIS overpass time. All of these factors may lead to satellite-based underestimation of fires and resulting emissions regardless of the fire detection algorithm strength in GFED or any other emission database. We further note that the GFED agricultural waste burning data includes emissions from all agricultural sources, yet the emissions are still notably lower than rice residue burning alone. However, these results are presented with the caveat that the field-based emission estimates could be improved through refinements mentioned in the limitations section. We also note for GFED emissions error rates that emission factor error is included, but not burned area error.

**Conclusion**

The first part of our study characterized total post-harvest rice residues for the small-holder rice-dominated province of Hanoi, Vietnam. From the field, we developed separate straw, stubble, and total biomass fuel-loading factors representative of a typical double-cropped rice region of Vietnam. Our results on rice residues suggested relatively higher total post-harvest residue factor than the earlier studies. We used Sentinel-1 C-band SAR data for field sampling of residues and to infer the relationship between post-harvest biomass and SAR. We found a moderately weak relationship between the SAR backscatter and field-measured rice straw, while a moderate relationship was observed with the total rice biomass. These results are promising, however, more advanced modelling might be necessary for forecasting the post-harvest biomass using SAR data. We note that not all field data was available for SAR modeling due to GPS or geolocation issues.

Using field data from multiple studies, we then estimate residue burning emissions. We found that rice residue burning accounts for about ∼13% of total PM$_{2.5}$ emissions in Vietnam, and is the second largest PM$_{2.5}$ source after fuelwood burning. We compared GFED-derived biomass burning emissions and crop waste burning emissions with our field-derived residue burning emissions. We found a likely notable underestimation by GFED for Vietnam by a factor of over 13. These results suggest a need for improved satellite-derived estimates. Further, emissions mitigation in this sector may be more critical than previously known and is increasingly important for health and air quality concerns in Hanoi, Vietnam.
Acknowledgments

Kristofer Lasko would like to thank his PhD committee members for guidance and support including Ivan Csizsar (NOAA), Matthews Hansen (UMd), and Louis Giglio (UMd). The authors would also like to thank Ha Van Pham (Vietnam National University Hanoi) for field assistance, David Lasko (Integrity Applications Incorporated) for verifying field calculations, as well as William Salas and Nathan Torbick from Applied Geosolutions LLC for SAR expertise. This research was funded in part by the Green Fellowship for the Environment, Council on the Environment, University of Maryland. We would also like to thank the anonymous referees for providing feedback which enhanced the quality of the manuscript.

References

Aalde H et al 2006 Generic methodologies applicable to multiple land-use categories IPCC Guidelines for National Greenhouse Gas Inventories vol 4 pp 1–59 (www.ipcc-nggip.iges.or.jp/public/2006gl/pdf/4_Volume4/V4.02.Ch2_Generic.pdf)

Aagi S K et al 2011 Emission factors for open and domestic biomass burning for use in atmospheric models Atmos. Chem. Phys. 11 4039–72

Arai H, Hosen Y, Pham Hong V N, Thi N T, Huu C N and Inubushi K 2015 Greenhouse gas emissions from rice straw burning and straw-mushroom cultivation in a triple rice cropping system in the Mekong Delta Soil Sci. Plant Nutr. 61 719–35

Badarinath K V S, Khalor S K and Sharma A R 2009 Long-range transport of aerosols from agriculture crop residue burning in Indo-Gangetic Plains—a study using LIDAR, ground measurements and satellite data J. Atmos. Solar Terr. Phys. 71 112–20

Badarinath K V S et al 2007 Multiyear ground-based and satellite observations of aerosol properties over a tropical urban area in India Atmos. Sci. Lett. 8 7–13

Badarinath K V S, Sharma A R, Khalor S K and Prasad V K 2009 Variations in CO, O3 and black carbon aerosol mass concentrations associated with planetary boundary layer (PBL) over tropical urban environment in India J. Atmos. Chem. 62 73–86

Biswas S, Vadrevu K P, Lwin Z M, Lasko K and Justice C O 2015 Factors controlling vegetation fires in protected and non-protected areas of Myanmar PLOS One 10 e0124346

Bouvet A et al 2014 Estimation of agricultural and biophysical parameters of rice fields in Vietnam using X-band dual-polarization SAR 2014 IEEE Geoscience and Remote Sensing Symp. 1504–7

Cao G, Zhang X, Suning and Zheng F 2008 Investigation on emission factors of particulate matter and gaseous pollutants from crop residue burning J. Environ. Sci. 20 50–5

Chakraborty M, Manjunath K R, Panigrahi S, Kundu N and Parihar I S 2005 Rice crop parameter retrieval using multi-temporal, multi-incidence angle Radarsat SAR data ISPRS J. Photo. Remote Sens. 59 310–22

Cristofanelli P et al 2014 Transport of short-lived climate forcers/ pollutants (SLCF/P) to the Himalayas during the South Asian summer monsoon onset Environ. Res. Lett. 9 084005

Czurda P J and Andreae M O 1990 Biomass burning in the tropics: Impact on atmospheric chemistry and biogeochemical cycles Science 250 1669–79

Dong J and Xiao X 2016 Evolution of regional to global paddy rice mapping methods: a review ISPRS J. Photo. Remote Sens. 119 214–27

Duong P T and Yoshio H 2015 Current situation and possibilities of rice straw management in Vietnam (www.jstnai.jp/Annual_Meeting/PROG_52/ResumeC/C02-4.pdf)

Erten E, Lopez-Sanchez J M, Yuzugullu O and Hajnsek I 2016 Retrieval of agricultural crop height from space: a comparison of SAR techniques Remote Sens. Environ. 187 130–44

Gadde B, Bonnet S, Menke C and Garvait S 2009 Air pollutant emissions from rice straw open field burning in India, Thailand and the Philippines Environ. Poll. 157 1554–8

Gadde B, Menke C and Wasmann R 2009 Rice straw as a renewable energy source in India, Thailand, and the Philippines: overall potential and limitations for energy contribution and greenhouse gas mitigation Biomass and Bioenergy 33 1532–46

Gebhardt S, Huth J, Lam-dao N, Roth A and Kuenzer C 2012 A comparison of TerraSAR-X Quadpol backscattering with RapidEye multispectral vegetation indices over rice fields in the Mekong Delta, Vietnam Int. J. Remote Sens. 33 7644–61

Giglio L, Randerson J T and van der Werf G R 2013 Analysis of daily, monthly, and annual burned area using the fourth-generation global fire emissions database (GFED4) J. Geophys. Res.: Biogeosci. 118 317–28

Gupta P K et al 2001 Study of trace gases and aerosol emissions due to biomass burning at shifting cultivation sites in East Godavari District (Andhra Pradesh) during INDOEX IFP-99curr. Sci. Ind. 80 186–96

Harvard University Department of Physics 2013 A Summary of Error Propagation (http://ipl.physics.harvard.edu/wp-uploads/2013/03/PS3_Error_Propagation_sp13.pdf)

Heinsch F A, McHugh C W and Hardy C C 2016 Fire, fuel, and smoke science program 2015 research accomplishments (Missoula, MT: US Department of Agriculture, Forest Service, Rocky Mountain Research Stations, Missoula Fire Sciences Laboratory) (www.firelab.org/sites/default/files/images/downloads/FIS_AnnualReport_FY2015.pdf)

Hien P D, Bac V T, Tham H C, Nhan D D and Vinh L D 2002 Influence of meteorological conditions on PM2.5 and PM2.5–10 concentrations during the monsoon season in Hanoi, Vietnam Atmos. Environ. 36 5473–84

Hong Van N P, Nga T T, Arai H, Hosen Y, Chiem N H and Inubushi K 2014 Rice straw management by farmers in a triple rice production system in the Mekong Delta, Vietnam Nnam. Trop. Agri. Dev. 58 155–62

Inoue Y, Sakaiya E and Wang C 2014 Capability of C-band backscattering coefficients from high-resolution satellite SAR sensors to assess biophysical variables in paddy rice Remote Sens. Environ. 140 257–66

Kanabkaew T and Oanh N T K 2011 Development of spatial and temporal emission inventory for crop residue field burning Environ. Mod. Asses. 16 453–64

Kanokklanjan K, Cheewapongprhan P and Garivati S 2011 Black carbon emission from paddy field open burning in Thailand IPCC BEEC Proc. 6 88–92

Karla K, Nevalainen O, Krooks A, Karjalainen M and Kaasalainen S 2014 Monitoring changes in rice cultivation area from SAR and optical satellite images in Ben Tre and Tra Vinh provinces in Mekong delta, Vietnam Remote Sens. 6 4090–108

Korenaga T, Liu X and Huang Z 2001 The influence of moisture content on polycyclic aromatic hydrocarbons emission during rice straw burning Chemosphere 3 117–22

Kuenzer C and Knauer K 2013 Remote sensing of rice crop areas Int. J. Remote Sens. 34 2101–39

Kurokawa J et al 2013 Emissions of air pollutants and greenhouse gases over Asian regions during 2000–2008: Regional Emission inventory in Asia (REAS) version 2. Atmos. Chem. Phys. 13 11019–58
Lam-Dao N, Le Toan T, Apan A, Bouvet A, Young F and Le-Van T 2009 Effects of changing rice cultural practices on C-band synthetic aperture radar backscatter using Envisat advanced synthetic aperture radar data in the Mekong River Delta J. Appl. Remote Sens. 3 033563
Le TH, Nguyen TN, Lasko K, Ilavajhala S, Vadrevu KP and Justice C 2014 Vegetation fires and air pollution in Vietnam Remote Sens. Pollut. 195 267–75
Liang S 2005 Quantitative Remote Sensing of Land Surfaces (New York: Wiley)
McNairn H and Brisco B 2004 The application of C-band polarimetric SAR for agriculture: a review Can. J. Remote Sens. 30 323–42
Mosleh M K, Hassan Q K and Chowdhury E H 2015 Application of remote sensors in mapping rice area and forecasting its production: a review Sensors 15 769–91
Nguyen M D 2012 An estimation of air pollutant emission from open rice straw burning in Red River Delta Sci. Dev. J. 10 190–98 (in Vietnamese)
Nguyen D, Wagner W, Nacimi V and Cao S 2015 Rice-plantated area extraction by time series analysis of ENVISAT ASAR WS data using a phenology-based classification approach: a case study for Red River Delta, Vietnam Int. Arch. Photogramm. 40 77
Nguyen T T et al 2015 Particulate matter concentration mapping from MODIS satellite data: a Vietnamese case study Environ. Res. Lett. 10 095016
Nguyen V H et al 2016 Energy efficiency, greenhouse gas emissions, and cost of rice straw collection in the Mekong River Delta of Vietnam Field Crops. Res. 198 16–22
Oanh et al 2011 Characterization of particulate matter emission from open burning of rice straw Atmos. Environ. 45 493–503
Orrit et al 2015 Regional diagnosis of biomass use in suburban village in southern Vietnam J. Japan Inst. Energy 94 805–29
Paloscia S, Macelloni G, Pampaloni P and Sigismondi S 1999 The potential of C-and L-band SAR in estimating vegetation biomass: the ERS-1 and JERS-1 experiments IEEE Trans. Geosci. Remote Sens. 37 2107–10
Patanonthai A 1996 Soils Under Stress: Nutrient Recycling and Agricultural Sustainability in the Red River Delta of Northern Vietnam (Hanoi, HE: East-West Center) (http://hdl.handle.net/10125/21608)
Pham V C, Pham T T H, Tong T H A and Pham N H 2015 The conversion of agricultural land in the peri-urban areas of Hanoi (Vietnam): patterns in space and time J. Land Sci. 10 224–42
Ponette-González A et al 2016 Biomass burning drives atmospheric nutrient redistribution within forested peatlands in Borneo Environ. Res. Lett. 11 085003
Rajput P, Sarin M, Sharma D and Singh D 2014 Characteristics and emission budget of carbonaceous species from post-harvest agricultural-waste burning in source region of the Indo-Gangetic Plain Tellus B 66 21026
Randerson J T et al 2012 Global burned area and biomass burning emissions from small fires J. Geophys. Res.: Biogeosci. 117 G4
Reddington C L et al 2014 Contribution of vegetation and peat fires to particulate air pollution in Southeast Asia Environ. Res. Lett. 9 094006
Ribbes F and Le Toan T 1999 Coupling radar data and rice growth model for yield estimation Geoscience and Remote Sensing Symp. IGARSS’99 Proc. 4 2336–8
Satalino G et al 2015 Retrieval of wheat biomass from multitemporal dual polarised SAR observations 2015 IEEE Int. Geoscience and Remote Sensing Symp. 5194–7
Streets D G et al 2004 On the future of carbonaceous aerosol emissions J. Geophys. Res.: Atmos. 109 D12
Streets D G, Yarber K F, Woo J H and Carmichael G R 2003 Biomass burning in Asia: annual and seasonal estimates and atmospheric emissions Glob. Biogeochem. Cycles 17 4
Sutherland W J 2006 Ecological Census Techniques: a Handbook vol 2 (Cambridge: Cambridge University Press)
Torbick N et al 2011 Integrating SAR and optical imagery for regional mapping of paddy rice attributes in the Poyang Lake Watershed, China Can. J. Remote Sens. 37 17–26
Torbick N, Salas W, Chowdhury D, Ingraham P and Trinh M 2017 Mapping rice greenhouse gas emissions in the Red River Delta, Vietnam Carbon Manage. 8 99–108
Trach N X 1998 The need for improved utilisation of rice straw as feed for ruminants in Vietnam: an overview Livestock Res. Rural Dev. 10 1–14
Tran S N, Nguyen T H N, Nguyen H C, Nguyen V C N, Le H V and Ingvorsen K 2014 UOC TINH LUONG VÀ CÁCH BIẾN PHÁP XỬ LÝ ROM RAO MỘT SỐ TINH ĐÔNG BÀNG SÔNG CƯƠI LONG Can. The Univ. Sci. J. 32 87–95
Vadrevu K P and Lasko K 2015 Satellite-derived nitrogen dioxide variations from biomass burning in a subtropical evergreen forest, northeast India Remote Sensing of Water Resources, Disasters, and Urban Studies (Boca Raton, FL: CRC Press) pp 545–59
Vadrevu K P, Elicott E, Badarinath K V S and Vermote E 2011 MODIS derived fire characteristics and aerosol optical depth variations during the agricultural residue burning season, North India Environ. Poll. 159 1560–9
Vadrevu K P et al 2012 Vegetation fires in the Himalayan region—aerosol load, black carbon emissions and smoke plume heights Atmos. Environ. 47 241–51
Vadrevu K P, Lasko K, Giglio L and Justice C 2014 Analysis of Southeast Asian pollution episode during June 2013 using satellite remote sensing datasets Environ. Poll. 195 245–56
Vadrevu K P, Lasko K, Giglio L and Justice C 2015 Vegetation fires, absorbing aerosols and smoke plume characteristics in diverse biomass burning regions of Asia Environ. Res. Lett. 10 105003
Van der Werf G et al 2010 Global fire emissions and the contribution of deforestation, savanna, forest, agricultural, and peat fires (1997–2009) Atmos. Chem. Phys. 10 11707–35
Vietnam Government 2015 General Statistics Office of Vietnam (www.gso.gov.vn/default_en.aspx)
Wilson A M and Walter J 2016 Remotely sensed high-resolution global cloud dynamics for predicting ecosystem and biodiversity distributions PLoS Biol. 14 e1002415
Wiseman G, McNairn H, Homayouni S and Shang J 2014 RADARSAT-2 Polarimetric SAR response to crop biomass for agricultural production monitoring IEEE J. Select. Topics Appl. Earth Observ. Remote Sens. 7 4461–71
Yan X, Ohara T and Akimoto H 2006 Bottom-up estimate of biomass burning in mainland China Environ. Res. 40 5262–73
Yevich R and Logan J A 2003 An assessment of biofuel use and burning of agricultural waste in the developing world Glob. Biogeochem. Cycles 17 4
Zhang T, Wang C, Wu J, Qi J and Salas W 2009 Mapping paddy rice with multitemporal ALOS/PALSAR imagery in southeast China Int. J. Remote Sens. 30 6301–15
Zhang T, Wooster M J, Green D C and Main B 2015 New field-based agricultural biomass burning trace gas, PM2.5, and black carbon emission ratios and factors measured in situ at crop residue fires in Eastern China Atmos. Environ. 121 22–34
Zuhlke M et al 2015 SNAP (Sentinel Application Platform) and the ESA Sentinel 3 Toolbox ESA Special Publication 734 21