Towards Sustainable Farm Production System: A Case Study of Corn Farming

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Abstract: Many recent studies show that most of the crop production systems in developing countries are not environmentally sustainable. This study uses the life cycle assessment (LCA) to investigate the potential impacts of corn production in Pakistan on global warming and human health damages and also suggests mitigation strategies to reduce environmental impacts towards sustainable crop production based on the results. Land-based, mass-based, and energy-based functional units were used. IMPACT 2002+ methodology—a combination of IMPACT 2002, Eco-Indicator 99, CML, and intergovernmental panel on climate change (IPCC)—is used for the impact assessment. The results demonstrated that the global warming potential of one-ton production of corn, one-hectare corn farm, and production of 1000 MJ energy were 354.18, 34,569.90, and 1275.13 kg CO$_2$ equivalents, respectively. The off-farm and on-farm emissions of nitrogen-based chemical fertilizers were the hotspots in the most impact categories. Moreover, human health damages followed by global warming as environmental externalities were also associated with corn production. We also highlighted the production areas with light, medium and extreme environmental externalities with Toba Tek Singh and Okara districts in the Punjab province of Pakistan being the most and least contributing districts towards global warming, respectively. Results further indicated that a 5 to 100% reduction of chemical fertilizers would mitigate the environmental impacts of corn production by 4.38 to 87.58% and 2.16 to 43.30% in terms of aquatic acidification and global warming, respectively. Modern farming systems and conservation technologies were suggested to reduce emissions and improve the environmental performance of corn production. Furthermore, agricultural extension and the ministry of agriculture should pay more attention to farmers’ education on emissions from farming inputs and their impact on climate.

Keywords: corn production; life cycle assessment; environmental impact; global warming; human health damages

1. Introduction

Climate change caused by the increasing concentration of greenhouse gases in the environment is a significant issue defying policymakers around the world. Major contributing factors include increased utilization of non-renewable energy resources since the industrial revolution in the eighteenth century and expanded farming activities associated with the growing population [1]. A growth rate of 4.1% in greenhouse gas emissions (GHG) was observed in Pakistan from 1994 to 2012, with energy and agriculture as the major stakeholders responsible for 89% of the total emissions [2]. As the agriculture sector is widely dispersed throughout the country and alone has contributed 44% to CO$_2$ equivalent...
GHG emissions [3], considerable mitigation strategies are required in this sector. Mitigation practices may include crop, land, and energy management, efficiency improvement measures, and provision of renewable energy systems [4]. Before suggesting a particular mitigation strategy to reduce environmental impacts, it is imperative to study the life cycle of crops [5].

The life cycle approach is considered to be the main tool for addressing the production and consumption challenges of agricultural systems. Life cycle-based approaches support the transition to improving the sustainability of existing production and consumption practices, e.g., LCA (life cycle assessment), SLCDA (Social Life Cycle Assessment), LCDA (Life Cycle Cost), and LEXA (Overall Life Cycle Sustainability Assessment). LCA is the most commonly used method for environmental impact assessment of a product over its life. Previous studies supported the application of LCA in agricultural production systems [6,7].

Moreover, in the agricultural domain LCA was used on various crops such as wheat [8], rice [9], maize [10], sugarcane [11], and cotton [12], as well as fruits and vegetables such as apples, sweet berries, plums [13,14], peaches [15], grapes [16], tomatoes [17], and carrots [18]. In addition to the stepwise examination of environmental burdens due to agricultural activities/inputs [19], the LCA can also be used to develop climate change adaptation strategies under existing production methods by scientists, managers, and decision-makers as a decision-making tool [20].

Pakistan is one of the top 20 corn-producing countries of the world [21]; corn itself is the third most important cereal crop in the country after rice and wheat. Due to its flexible domain as a grain and cash crop, it is quickly becoming the main crop in Pakistan [22]. As a vital crop, the utilization of inputs has been increased as farmers presumed that increased application would lead to increased yield and eventually increased farm profit. Climate change and greenhouse gas emissions from crop productions are mainly due to the excessive use of inputs especially fossil fuels, synthetic fertilizers, and pesticides [23].

Agriculture, however, has the potential to mitigate its emissions through the adoption of appropriate measures like low carbon and resource conservation [24]. Therefore, an in-depth understanding of the trends in emissions of greenhouse gases, drivers of these emissions, and the linkage between them is essential to develop mitigation and adaptation strategies. In this regard, life cycle assessment (LCA) as a standard-based methodology that is accepted by researchers, policymakers, and industries can play a key role in a sustainable agricultural production system [25].

Despite the importance of this particular food crop (corn) and its bulky annual global production of 1116.41 million tons [21], studies of its environmental performance are still rare in the literature. The lack of life cycle assessment studies of corn and knowledge about environmental hotspots in the corn production chain is an important research gap. There are many studies on the LCA of crops such as wheat, rice, sugarcane, and cotton. However, due to specific crop inputs requirements and climatology, these studies could not be used as a benchmark for other crops in different continents. Furthermore, the ecological performance of a crop could also diverge from the economic status and industrial progress of production farms, and therefore, it is imperative to apply the LCA in different areal domains to establish production benchmarks for individual crops.

There have been a few studies of the LCA applications on Pakistani agriculture [26,27]. However, to the best of the authors’ knowledge, no study exists regarding the environmental externalities of corn productions in the country. Therefore, this study emphasizes the LCA of corn production to investigate environmental impacts and improvement options. The aggregated, normalized, and weighted impact assessment measures are established for resource consumption and land uses. Moreover, currently, the best accessible techniques have been selected and refined to incorporate standardizations and weighting factors for impact categories. The current study described the LCA application on Pakistani corn with specific emphasis on the identification of hotspots in the production chain. The objectives of this study were to: (1) identify the LCA impacts of Pakistani corn and environmen-
tal hotspots in corn farming; (2) quantitatively analyze the impact of corn growing on human health, ecosystem, climate change, and resources and explore the mechanism of climate change and human activities on the production chain; and (3) provide the available mitigation options for corn production in Pakistan.

2. Materials and Methods

The LCA is divided into four steps [28].

- “Definition of the goal and scope”;
- “Life cycle inventory”;
- “Life cycle impact assessment”;
- “Interpretation of results”.

2.1. First Step—Definition of the Goal and Scope

2.1.1. Goal and Scope

The goal of this research is to determine the environmental LCA of corn production in the Punjab province of Pakistan. Three functional units (FUs) namely, land-based (one hectare), mass-based (one tonne of corn), and energy-based (1000 MJ energy generation) units were considered.

2.1.2. Case Study

The case study was corn production system in the Punjab province of Pakistan, a key focus of the study. Figure 1 shows the study area with coordinates of potential districts of corn production in Punjab.

The data were collected from 500 corn producers through a well-structured questionnaire from 2018 to 2019. Following Elahi et al. [29,30], the questionnaire was developed to conduct interviews with farmers of Punjab. Each producer was asked questions of the number of inputs used from cultivation to harvesting such as diesel fuel (liters), machinery use (h), chemical fertilizers (kgs), biocides (kgs), seed (kgs), and yield (kgs).

The selection of these farmers has been made through a five-stage sampling technique. In the first stage, Punjab province was chosen purposively since it is the major corn-producing province in the country. Secondly, 10 municipalities (MP) of the corn-rice...
cropping zone were selected. In the third and fourth stages, 2 union councils (UCs) and 5 villages were randomly selected. Finally, 5 farmers from each V were chosen for the interview. Therefore, 50 farmers from every MP were interviewed for inputs and output of corn productions. The questionnaire was designed to obtain per acre information (a unit of land measurement locally used in Pakistan) converted to per hectare.

For the scope of LCA in corn productions, cradle to farm gate system boundary was used. The emitted pollutants were divided into two groups of off-farm emissions (i.e., the emitted pollutants from extraction, production, and transportation of raw materials), and on-farm emissions (i.e., direct emissions which were released at the farm).

2.2. Second Step—Life Cycle Inventory

In this step, the input consumption and their emitted pollutants were investigated. Within the system boundary of the cradle-to-farm gate, both foreground and background emissions from corn productions were considered. Background emissions include emission from production (at industry) and transportation of material inputs (to farm), while foreground includes the direct emission from consumption of materials inputs at the farm. The data collected through questionnaires described in the previous section were used for foreground emissions, while the associated background emissions were extracted from EcoInvent 2.2 database [31]. For foreground emissions, the quantity of each input was multiplied by its CO$_2$ equivalent, and the LCA included its algorithms to evaluate these equivalent CO$_2$ emissions of any processes throughout the corn production cycle.

2.3. Third Step—Life Cycle Impact Assessment

The impact assessment (IA) methodology was used. In this methodology selection of impact categories and characterization are mandatory, while normalization and weighing are optional elements [32]. We selected fifteen impact categories for investigation. The characterization index was calculated by using the potential of pollutants in these impact categories. Then, the values were normalized, the normalization indices were calculated by dividing the characterization indices by normalization factors based on the following equation:

$$NI_i = \frac{CI_i}{NF_i}$$  \hspace{1cm} (1)

where, $NI$, $CI$, and $NF$ are the normalization index, characterization index, and normalization factor of impact category $i$, respectively [33].

For IA of LCA, IMPACT 2002+ methodology—a combination of IMPACT 2002, CML, IPCC, and Eco-Indicator 99—was used in this study [34]. The IA methodology is strongly based on the initial outcomes from life cycle inventory (LCA), the calculated emissions in the LCI results were then entered into the intermediate stage, where the results with similar impact pathways were grouped into impact categories called midpoint categories. Later on, these results of the midpoint were then quantified to represent the quality change in terms of four damage categories at the endpoint level. The overall framework of impact methodology has been represented in Figure 2.

Since different studies used different methods for impact IA in crop productions, a comparative analysis was carried out to examine the effects of different IA methods on the final results. To do so, four different methods of IA methods of CML-IA baseline, ILCD 2011 Midpoint, EDIP 2003, and EDP (2013) were evaluated for comparison through LCA with SimaPro 8.2.3.0 version [32].
Figure 2. A systemic scheme of the IA methodology of corn production.

3. Results and Discussions

3.1. Step Four—Interpretation of LCA Results

This step involves the interpretation of the IA findings. The study considered 15 IA categories. The characterization indices of these categories for corn production are presented in Table 1. Impact categories like global warming potential (GWP) are considered as imperious concerning the ecological recitation of the crops [35]. Results showed that field emission and nitrogen fertilizers are the primary contributors to the GWP of the corn cropping system. Field emission is the main source of eutrophication, aquatic acidification, and CO$_2$. The application of nitrogen fertilizer is responsible for nitrous oxide (N$_2$O), which is 310 times more dangerous than CO$_2$ due to its heat-trapping effects [36]. Mass (one tonne), land (one hectare), and energy (1000 MJ) based GWP of corn production was found 354.18, 34,569.90, and 1275.13 kg CO$_2$ eq, respectively. In the literature, mass-based GWP of canola production was reported to be 1181 kg CO$_2$ eq [37]; traditional and semi-mechanized potato to be 228 and 153 kg CO$_2$ eq, respectively [38]; canola, soybean, and sunflower to be 2132, 1549, 2283 kg CO$_2$ eq, respectively [39]; and irrigated and dryland wheat to be 1220 and 720, respectively [40].

Therefore, compared to these studies the GWP of corn production in Pakistan is not high. Moreover, the amount of eutrophication and acidification indices are almost the same as the corresponding values in different study regions. The results indicate that corn production is inefficient in terms of fuel and fertilizer application, which contributes the most towards global warming.

Table 2, illustrates the percentage involvement of each input towards impact categories. The results emphasized that direct emissions were the hotspot in terms of aquatic acidification, terrestrial acid/nutria, and respiratory inorganics. Mousavi-Avval et al. [37] reported that 70% of the environmental impact of terrestrial ecotoxicity within corn production was related to on-farm emissions. Ghasempour and Ahmadi [41] studied a corn production system in Iran and estimated that nitrogen-based fertilizers caused the highest negative impacts. Heidari et al. [42] studied the LCA of pasta production and found that foreground emissions were the major stakeholders towards environmental externalities.

Furthermore, nitrogen-based fertilizers were the flashpoint in terms of GWP, mineral extraction, non-renewable energy, terrestrial ecotoxicity, and carcinogens. The application of farm inputs especially fertilizers was different at different farms depending upon farmers’ specific characteristics, for example, their knowledge about precision agriculture, financial condition, extension services, soil fertility, and crop rotation. Joy et al. [43] also reported an imbalanced use of fertilizers in Pakistan. As the excessive use of fertilizers caused more nitrogen emission, and thus, increased global warming, farmers should pay special attention to the pattern of use of fertilizers. Another important non-renewable energy...
input is the diesel fuel use in different machines such as tractors for crop activities (from land preparation to harvesting) and pumps for groundwater pumping, which also causes on-farm emissions.

Table 1. Characterization indices of the corn production system.

| Impact Categories          | Units                                | Functional Units                  |
|----------------------------|--------------------------------------|-----------------------------------|
|                            | Land-Based                          | Mass-Based                        | Per 1000 MJ Energy Generation |
| Carcinogens                | Kg C\textsubscript{2}H\textsubscript{3}Cl equivalent | 87.24                             | 13.34                         | 48.01                          |
| Non-carcinogens            | Kg C\textsubscript{2}H\textsubscript{3}Cl equivalent | 48.67                             | 7.36                          | 26.48                          |
| Respiratory inorganics     | Kg PM\textsubscript{2.5} equivalent   | 8.56                              | 1.31                          | 4.71                           |
| Ionizing radiation         | Bq C-14 equivalent                  | 20,111.04                         | 3087.28                       | 11,151.61                      |
| Ozone layer depletion      | Kg CFC-11 equivalent                | 6.40 × 10\textsuperscript{-4}      | 9.99 × 10\textsuperscript{-5}  | 0.00036                        |
| Respiratory organics       | Kg C\textsubscript{2}H\textsubscript{4} equivalent | 0.84                              | 0.13                          | 0.47                           |
| Aquatic ecotoxicity        | Kg TEG water                        | 231,588.27                        | 35,152.98                     | 126,571.60                     |
| Terrestrial Ecotoxicity    | Kg TEG soil                         | 29,304.28                         | 4483.80                       | 16,145.65                      |
| Terrestrial acid/nutria    | Kg SO\textsubscript{2} equivalent    | 742.12                            | 113.38                        | 408.17                         |
| Land occupation            | m\textsuperscript{2} arable         | 97.90                             | 14.83                         | 53.42                          |
| Aquatic acidification      | Kg SO\textsubscript{2} equivalent    | 102.78                            | 15.70                         | 56.53                          |
| Aquatic eutrophication     | Kg PO\textsubscript{4} P-lim        | 1.39                              | 0.21                          | 0.76                           |
| Global warming             | Kg CO\textsubscript{2} equivalent   | 2295.34                           | 354.18                        | 1275.13                        |
| Non-renewable energy       | MJ primary                          | 34,569.90                         | 5535.04                       | 19,272.15                      |
| Mineral extraction         | MJ surplus                          | 165.01                            | 25.05                         | 90.18                          |

Note: C\textsubscript{2}H\textsubscript{3}Cl Vinyl chloride; PM Particulate matter; CFC chlorofluorocarbons; TEG Triethylene glycol; C\textsubscript{2}H\textsubscript{4} Ethylene; SO\textsubscript{2} Sulphur dioxide; PO\textsubscript{4} Phosphate.

Table 2. Percentage share of inputs towards impact categories.

| Impact Categories          | Seed | Potassium Fertilizer | Phosphate Fertilizer | Nitrogen Fertilizer | Pesticide | Diesel Fuel | On-Farm |
|----------------------------|------|----------------------|----------------------|--------------------|-----------|-------------|---------|
| Carcinogens                | 0.9  | 0.9                  | 14.0                 | 64.5               | 16.0      | 3.7         | 0.0     |
| Non-carcinogens            | 4.8  | 5.2                  | 39.4                 | 31.0               | 13.5      | 1.6         | 4.5     |
| Respiratory inorganics     | 0.0  | 1.0                  | 10.4                 | 11.9               | 5.6       | 1.0         | 70.1    |
| Ionizing radiation         | 1.6  | 1.1                  | 34.7                 | 23.3               | 22.0      | 17.3        | 0.0     |
| Ozone layer depletion      | 0.5  | 0.0                  | 3.2                  | 5.4                | 84.4      | 6.5         | 0.0     |
| Respiratory organics       | 1.0  | 2.0                  | 12.4                 | 33.4               | 21.0      | 30.2        | 0.0     |
| Aquatic ecotoxicity        | 2.9  | 1.1                  | 53.3                 | 22.1               | 17.6      | 3.0         | 0.0     |
| Terrestrial Ecotoxicity    | 9.0  | 9.0                  | 30.0                 | 36.0               | 17.7      | 5.0         | 1.3     |
| Terrestrial acid/nutria    | 0.7  | 0.0                  | 2.3                  | 2.7                | 1.3       | 0.0         | 93.0    |
| Land occupation            | 39.0 | 15.0                 | 37.0                 | 5.0                | 4.0       | 0.0         | 0.0     |
| Aquatic acidification      | 0.0  | 3.0                  | 5.0                  | 6.0                | 3.0       | 0.0         | 85.0    |
| Aquatic eutrophication     | 0.0  | 1.5                  | 64.5                 | 11.8               | 20.2      | 2.0         | 0.0     |
| Global warming             | 1.1  | 2.5                  | 15.0                 | 45.0               | 12.0      | 3.4         | 21.0    |
| Non-renewable energy       | 1.0  | 1.4                  | 15.0                 | 36.4               | 17.6      | 28.6        | 0.0     |
| Mineral extraction         | 1.0  | 1.0                  | 39.0                 | 46.0               | 13.0      | 0.0         | 0.0     |

Table 3, further demonstrates the normalized damage assessment (NDA) of the corn crop. Human health stands for the highest damaged category caused by direct emissions due to the application of fertilizers. The results are in line with the study of Elahi et al. [44]. They have also found associated occupational health damages due to the application of synthetic agrochemicals for crop production in the Punjab province of Pakistan. Furthermore, nitrogen fertilizers have also the highest negative effect on the resource depletion damage category followed by diesel fuel consumption. Our results are in line with the studies of Payandeh et al. [45] and Taki et al. [46].

Table 3. Normalized damage for different input consumptions.

| Damage Categories          | Seed | Potassium Fertilizer | Phosphate Fertilizer | Nitrogen Fertilizer | Pesticide | Diesel Fuel | On-Farm |
|----------------------------|------|----------------------|----------------------|--------------------|-----------|-------------|---------|
| Human health               | 0.0045 | 0.0054             | 0.0989               | 0.1277             | 0.0551    | 0.0115      | 0.5964  |
| Eco systems                | 0.0046 | 0.0014             | 0.0095               | 0.0087             | 0.0039    | 0.0011      | 0.0527  |
| Climate change             | 0.0025 | 0.0028             | 0.0373               | 0.1002             | 0.0316    | 0.0094      | 0.0481  |
| Resources                  | 0.0018 | 0.0018             | 0.0341               | 0.0860             | 0.0389    | 0.0060      | 0.0000  |
Furthermore, the description of corn production indices through different impact assessment methodologies is presented in Table 4. Results indicate that the land-based GWP of corn ranges from 2295.34 kg CO₂ eq to 3134.44 kg CO₂ eq per hectare under different IA methodologies, while OLP (ozone layer depletion) stays the same at $6.40 \times 10^{-4}$ kg CFC-11 eq. for all IA methodologies. Rafiee et al. [47] also signified the variations of LCA results among the IA methodologies in the agriculture sector.

Table 4. Description of production indices calculated by various IA methods.

| IA                  | GWP (Kg CO₂ eq) | OLP (Kg CFC-11 eq) |
|---------------------|-----------------|--------------------|
| CML-IA baseline     | 3134.38         | $6.40 \times 10^{-4}$ |
| EDIP 2003           | 3134.44         | $6.40 \times 10^{-4}$ |
| EDP (2013)          | 3134.38         | $6.40 \times 10^{-4}$ |
| ILCD 2011 Midpoint  | 3135.25         | $6.40 \times 10^{-4}$ |
| IMPACT 2002+ *      | 2295.34         | $6.40 \times 10^{-4}$ |

*This IA method was considered in this study.

Figure 3 illustrates the NDA of the corn crop. Our results show that corn production in Pakistan causes the highest damage to the human health category. The results are in line with the study of Paramesh et al. [32]. They have estimated human health damage assessment as the largest environmental impact during areca nut production. The direct emissions into the air during the agricultural production process are most responsible for human health damages. Furthermore, climate change/global warming and resource depletion are found the second and third most damaging categories after human health. This is because a frequent number of irrigations and excessive input requirements for corn production contribute to climate change and resource depletion.

3.2. Distribution of Study Area Based on Impact Categories

Since the LCA results reveal nitrogen fertilizers as the main contributors towards global warming evidenced by the weighted global warming index of 2295.34 kg CO₂ equivalent, the utilization of nitrogen fertilizers was calculated for each district and the entire study region (10 districts) and then the districts were categorized into light, medium and extreme areas, shown in Figure 4. Light, medium, and extreme emission areas include the global warming index ranges from 1698 to 2019 kg CO₂ equivalent, 2020 to 2295 kg CO₂ equivalent, and above 2295 kg CO₂ equivalent, respectively. Okara, and Faisalabad, and Toba Tek Singh and Jhang are found to be the light and extreme regions, respectively. More
specifically, the farmers of Jhang and Toba Tek Singh have high use of nitrogen fertilizers. Previous studies [48–50] also concluded that inadequate usage of synthetic fertilizers was significantly related to emission and other related consequences. Zulfiqar and Thapa [51] determined the imbalanced use of fertilizers was the main reason for getting lower wheat efficiency in the Punjab province of Pakistan. Developing countries like Pakistan, where farmers expect to have increased crop yield with increased use of chemical fertilizers are becoming the major reasons for greenhouse gas emissions and climate change.

Figure 3. NDA of corn production.

Figure 4. Study area with light, medium, and extreme externalities.

3.3. Mitigation Strategies

Table 5 represents the potential of environmental impacts mitigation through the reduction of nitrogen-based chemical fertilizers. Results indicate that a 5 to 100% reduction in chemical fertilizers mitigates the environmental impacts of corn production in terms of aquatic acidification by 4.38 to 87.58%. It can be concluded that nitrogen-based fertilizer is an environmental hotspot in the corn production system. Sadeghi et al. [52] evaluated the environmental impacts of corn production in Iran based on the application of three nitrogen-based fertilizer levels of 300, 400, and 500 kg per hectare. The results showed that the corn yield increased with increasing fertilizer rate from 300 to 400 kg per hectare; however, a further increase in the fertilizer rate (500 kg ha\(^{-1}\)) was not environmentally friendly.

Nitrogen-based fertilizer is one of the main inputs in corn production [53,54]. Chemical fertilizer has been considered the main driver or input to increase crop production in developing countries such as the production of groundnut-bean [55], peach [15], and areca nut [32]. Integrating a legume into the crop rotation can contribute to the supply of required chemical nitrogen by atmospheric nitrogen fixation [37]. Moreover, planting winter cover crops can be considered as a strategy to increase soil organic carbon levels and to reduce nitrogen losses from the soil and to increase soil organic carbon levels [56,57] suggested that replacing conventional chemical fertilizers with biofertilizers could mitigate the environmental impacts of the crop production system. Currently, there is a common misconception among Pakistani farmers that the excessive use of fertilizers would increase crop yield. However, E. Elahi, Weijun, Zhang et al. [50] found that only an optimal use of NPK fertilizers could improve the crop efficiency.

Moreover, the excessive use of nitrogen-based fertilizers and fuel reduces energy use efficiency and threatens the environment. Prabhakar and Elder [58] also reported that most farmers in Pakistan did not use the optimal amount of farm inputs, and thus, resulted in low resource use efficiency. Zulfiqar and Thapa [51] found a lack of access to extension services, crop production technology, and knowledge about environmental impacts of crop production were among the main reason for low crop efficiency and negative environmental impacts.
Table 5. The environmental impact mitigation through the reduction of nitrogen-based fertilizers.

| Impact Categories       | Units                  | The Characterization Based on the Various Scenarios |
|-------------------------|------------------------|-----------------------------------------------------|
|                         | Current                | 5% N-Based Fertilizer Reduction | 10% N-Based Fertilizer Reduction | 20% N-Based Fertilizer Reduction | 50% N-Based Fertilizer Reduction | 100% N-Based Fertilizer Reduction |
| Carcinogens             | Kg C₂H₃Cl equivalent  | 87.24                               | 84.42                               | 81.58                               | 75.92                               | 58.93                               | 30.61                               |
| Non-carcinogens         | Kg C₂H₃Cl equivalent  | 48.67                               | 47.81                               | 46.94                               | 45.20                               | 39.99                               | 31.31                               |
| Respiratory inorganics  | Kg PM₂.₅ equivalent    | 8.56                                | 8.24                                | 7.91                                | 7.26                                | 5.31                                | 2.06                                |
| Ionizing radiation      | Bq C-14 equivalent     | 20,111.04                           | 19,879.33                           | 19,647.34                           | 19,183.56                           | 17,792.22                           | 15,473.33                           |
| Ozon layer depletion    | Kg CFC-11 equivalent   | 6.40 × 10⁻⁴                         | 6.39 × 10⁻⁴                         | 6.38 × 10⁻⁴                         | 6.34 × 10⁻⁴                         | 6.24 × 10⁻⁴                         | 6.07 × 10⁻⁴                         |
| Respiratory organics    | Kg C₂H₄ equivalent     | 0.84                                | 0.83                                | 0.82                                | 0.79                                | 0.70                                | 0.56                                |
| Aquatic ecotoxicity     | Kg TEG water           | 231,588.27                          | 229,090.20                          | 226,589.10                          | 221,589.20                          | 206,589.50                          | 181,589.90                          |
| Terrestrial Ecotoxicity | Kg TEG soil            | 29,304.28                           | 28,749.87                           | 28,209.10                           | 27,085.20                           | 23,756.33                           | 18,208.22                           |
| Terrestrial acid/nutria | Kg SO₂ equivalent      | 742.12                              | 707.41                              | 672.70                              | 603.28                              | 395.02                              | 47.92                               |
| Land occupation         | m² org.arable          | 97.90                               | 97.64                               | 97.36                               | 96.82                               | 95.18                               | 92.45                               |
| Aquatic acidification   | Kg SO₂ equivalent      | 102.78                              | 98.28                               | 93.77                               | 84.77                               | 57.77                               | 12.76                               |
| Aquatic eutrophication  | Kg PO₄ P-lim           | 1.39                                | 1.38                                | 1.37                                | 1.36                                | 1.31                                | 1.24                                |
| Global warming          | Kg CO₂ equivalent      | 2295.34                             | 2245.79                             | 2196.18                             | 2097.01                             | 1799.50                             | 1303.66                             |
| Non-renewable energy    | MJ primary             | 34,569.90                           | 33,920.62                           | 33,270.57                           | 31,971.05                           | 28,072.47                           | 21,574.85                           |
| Mineral extraction      | MJ surplus             | 165.01                              | 161.24                              | 157.46                              | 149.9                               | 127.23                              | 89.45                               |
4. Conclusions

This study applied the life cycle assessment (LCA) to determine the damage categories of global warming potential GWP, human health, resources depletion, and ecosystems of corn production in Pakistan. Human health was found to be the most damaged category, followed by climate change/global warming and resource depletion, and the direct emissions into the air during the corn production intercultural process are most responsible to the human health damages. On the other hand, frequent irrigations and excessive input consumptions for corn production contribute to climate change and resource depletion. The results further revealed that direct and indirect emissions of nitrogenous fertilizers were the environmental hotspot. The results also suggested that the reduction in the excessive use of farm inputs could be a potential sustainable crop production strategy. Therefore, the study calls for the proper use of agrochemicals. The on-farm emissions through fuel consumption can be minimized using conservation tillage, high-efficiency irrigation systems, and other integrated methods to control weeds and pests. Moreover, agricultural extension and the ministry of agriculture need to pay more attention to farmers’ education to improve energy resource use efficiency and reduction of negative environmental impacts of crop production. Further research is required to determine life cycle assessment in other agricultural production systems particularly focusing on climate change mitigation potential of different technologies and management practices.

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