BRIEF COMMUNICATION

The short-term effects of planktivorous fish foraging in the presence of artificial light at night on lake zooplankton

JOANNA TAŁANDA, PIOTR MASZCZYK*, EWA BABKIEWICZ, KATARZyna RUTKOWSKA AND MIROSŁAW ŚLUSARCZYK
DEPARTMENT OF HYDROBIOLOGY, INSTITUTE OF FUNCTIONAL BIOLOGY AND ECOLOGY, FACULTY OF BIOLOGY, UNIVERSITY OF WARSAW AT BIOLOGICAL AND CHEMICAL RESEARCH CENTRE, ŻWIRKI WIGURY 101, 02-089 WARSAW, POLAND

*CORRESPONDING AUTHOR: p.maszczyk@uw.edu.pl

Received April 11, 2022; revised July 14, 2022; accepted August 7, 2022

Corresponding editor: Marja Koski

Numerous studies have revealed that artificial light at night alters the natural patterns of light in space and time and may have various ecological impacts at different ecological levels. However, only a few studies have assessed its effect on interactions between organisms in aquatic environments, including predator–prey interactions in lakes. To fill this gap, we performed a preliminary enclosure experiment in which we compared the foraging effect of juvenile perch (Perca fluviatilis) on a natural lake zooplankton community in the absence and presence of light of high-pressure sodium (HPS) lamps mimicking artificial light emitted by a boat. The results revealed that even short-lasting exposure to HPS lamps may result in increasing fish predation, which in turn decreased the mean body size in zooplankton populations (e.g. Bosmina thersites) and affected the relative proportion between different taxa in zooplankton communities.

KEYWORDS: ALAN; community structure; light pollution; size-selective predation; size structure; zooplankton
INTRODUCTION

Artificial light at night (ALAN) alters the natural patterns of light in space and time, disrupting natural resource use and information flow. This may have ecological impacts at the individual, population and community levels (e.g. Longcore and Rich, 2004; Sanders et al., 2021; Zapata et al., 2019). Studies on the effects of ALAN at the population and community levels have been focused on interactions between organisms, including relations between a predator and its prey (e.g. Czarnecka et al., 2019; Fermin and Seronay, 1997; Nguyen and Winger, 2019; Nuñez et al., 2021). However, these effects in relation to the effects at the individual level remain relatively understudied (Gaston et al., 2015; Nuñez et al., 2021; Owens and Lewis, 2018; Sanders and Gaston, 2018). This issue is poorly explored even for planktivorous fish and zooplankton interactions—the best studied model of predator–prey interplay.

In lakes inhabited by fish, the foraging activity of visually oriented positive-size selective fishes is the main driver shaping space distribution, population size, age structure and community composition of planktonic animals (e.g. Gliwicz, 2003). ALAN by disrupting natural light intensity and spectral composition could alter the impact of fish predation on zooplankton and in turn on their population and community structure. However, the literature provides mainly indirect evidence supporting this prediction. ALAN was shown to attract fish (e.g. Becker et al., 2013; Nguyen and Winger, 2019) and enhance their foraging efficiency (Bolton et al., 2017; Fermin and Seronay, 1997). From the zooplankton perspective, it has been revealed that its response to the mortality risk in the presence of ALAN may be maladaptive (Maszczyk et al., 2021; Tałanda et al., 2018). Direct evidence concerns only the ALAN-dependent effects of fish predation on depth selection behavior of zooplankton populations and communities (Fermin and Seronay, 1997; Maszczyk et al., 2021; Moore et al., 2000; Springer and Skrzypczak, 2015), but neither on the mean body size nor community composition. The aim of our study was to fill this gap by assessing the short-term ALAN-dependent effect of a foraging planktivorous fish lake zooplankton body size and community composition in a preliminary enclosure experiment. More specifically, we hypothesized that the presence of artificial light produced by the high-pressure sodium (HPS) lamp may affect zooplankton vertical distribution and may increase positive size selective fish predation, which in turn decreases both the mean body size in zooplankton populations and the relative density of more vulnerable (i.e. larger-bodied and less evasive) species in zooplankton communities.

METHODS

The experiment was conducted on July 22–23, 2017 during the new moon in a small and shallow (7.0 m) bay located far from ALAN sources in the thermally stratified (Fig. 1), eutrophic (chlorophyll a conc. = 8.31 μg L⁻¹, Secchi disc = 2.6 m) Lake Roš (53°40’24.4”N 21°53’32.7”E; Great Mazurian Lakes; NE Poland). At the time of the experiment, the zooplankton community was dominated in decreasing order by *Cyclopoidea*, *Calanoida*, *Daphnia cucullata* and *Bosmina thersites*. The remaining taxa were less represented by at least an order of magnitude. The enclosures were made of transparent PVC (H = 7.7 m, Φ = 0.5 m), open at the top and closed at the bottom, and were attached to an anchored floating platform with an HPS artificial light source (70 W OSRAM®). The light source was attached to a non-transparent lampshade set at 2.4 m above water, mimicking the light emitted by a boat on a lake and producing approximately 8× greater light intensity and an entirely different spectral composition (e.g. Maszczyk et al., 2021) than full moonlight (Appendix 1). Three enclosures were placed on the “bright” side of the platform, while the 3 others were covered at the top with black non-transparent tarpaulin and placed on the opposite “dark” side. We used the zooplankton community originating as prey (collected from the whole water column in the experimental bay with a plankton net of 150-μm mesh size), and juvenile 0+ individuals of visually oriented planktivore fish *Perca fluviatilis* as a predator, all originating from Lake Roš. The fish were kept without food for 48 h before the experiment.

The day before the experiment, the enclosures were filled with filtered (5-μm mesh size) lake water (pumped by a petrol-driven water pump from a 1-m depth) and left for 5 hours to establish a natural temperature gradient. The zooplankton was introduced from the lake into the enclosures in natural densities at 7 p.m. At 12 p.m., vertical gradients of physical and chemical parameters of the water were measured in the enclosures (Fig. 1), and non-stratified samples of zooplankton from each of the enclosures were collected (integrated vertical sample). After that, the “dark” enclosures were covered with a tarpaulin and the light was switched on and then measured in each enclosure (Fig. 1). The experiment started at 1 a.m. (three hours after sunset) by adding 20 perch to each of the enclosures. At 4 a.m. (dawn), zooplankton stratified samples (0–2, 2–4 and 4–6 m) were collected from each enclosure, the light was switched off and then all fish were caught, euthanized, preserved in ethanol and transported to the lab, where their gastrointestinal tracts were dissected.
Zooplankton samples (initial and final) and fish gastrointestinal tracts from each of the enclosures ($n = 16.3 \pm 2.1$ and $17.3 \pm 1.5$ in dark and ALAN treatments) were analyzed under a stereomicroscope by counting and assigning individuals to six of the most numerous taxa (cladocerans: $B$. thersites, $D$. cucullata, $L$. kindti, $D$. brachyurum and copepods: Cyclopoida sp., Calanoida sp.). The length of 30 randomly selected individuals of $B$. thersites individuals (the most common taxa in the fish intestines) was also measured in each zooplankton sample. Statistical comparisons of the measured parameters between the two tested treatments were performed with a bootstrap approach that was used to simulate the distributions of the analyzed variables with the use of the collected data. The 95% bias-corrected confidence intervals were calculated for the differences between the illuminated and dark enclosures. Analyses of some parameters (e.g. taxon densities, body size of $B$. thersites) were performed on relative values considering the initial and final data in each enclosure. More details concerning the statistical procedures can be found in the Appendix 2.

RESULTS AND DISCUSSION

The data for all of the six taxa assessed together revealed that zooplankton resided closer to the water surface in illuminated enclosures compared with the dark ones (Fig. 1, Table 1 in the Appendix 1), which is in the line with several previous studies (e.g. Fermin and Seronay, 1997; Springer and Skrzypczak, 2015; Szczerbowski and Manczcz, 1984), but opposite to some others, which revealed a negative effect of ALAN on the depth distribution of zooplankton (Ludvigsen et al., 2018; Maszczyk et al., 2021; Moore et al., 2000; Rudstam et al., 1992). The observed variability could be attributed to differences in the light beam intensity, focus and variability (directed towards the water surface in the former studies or diffused and temporary variable in the latter ones). The positive phototaxis in our study may most likely be due to the maladaptive response of $D$. longispina perceiving the higher concentration of artificial light as a hint of the location of higher food concentration or temperature. More detailed analysis of our results revealed positive phototaxis to ALAN only in the case of smaller-bodied zooplankton taxa such as $B$. thersites, $D$. brachyurum and more evasive Cyclopoida, which are less vulnerable to positive size selective fish predation, but revealed negative phototaxis in the case of $D$. longispina and no effect in the case of remaining taxa ($L$. kindti and Calanoida). This may suggest that the depth selection behavior of zooplankton depends also on the ALAN-dependent perceived mortality risk from predation. It should be pointed out that even when some taxa exhibit negative phototaxis to ALAN, their response may be too weak in relation to the real light-dependent mortality risk and in turn be maladaptive (Maszczyk et al., 2021).

The number of each of the four taxa of cladocerans, their sum and the sum of all of the six analyzed taxa was greater in the guts of fish inhabiting illuminated enclosures compared with the dark ones (Fig 2a, Table 2 in the Appendix 1), which may be attributed to both attraction of zooplankton to the ALAN and their greater visibility in the presence of ALAN and in turn enhanced predation by visually oriented fish (Fermin and Seronay, 1997). Contrarily, the effect of ALAN on the number of eaten Cyclopoida and Calanoida was negligible, which may be due to their greater evasiveness in relation to the remaining zooplankton taxa.
While the final population density appears lower in all six analyzed taxa in the ALAN treatment than in the dark (Fig 2b), a significant difference was evidenced in the case of *Calanoida* and *D. brachyurum* merely (Table 3 in the Appendix 1), most likely due to the small number of replicates in the experiment. Therefore, the results confirm our predictions that the presence of ALAN may affect the zooplankton community composition but did not confirm that its presence may result in a relative decrease in the density of more vulnerable, i.e. larger-bodied and less evasive species (such as *D. longispina* and *L. kindti*).

The mean body length of the *B. thersites* population decreased during the experiment in both treatments, yet the difference was more apparent in the illuminated enclosures than in the dark ones (bootstrap distribution, mean bootstrap diff. = −0.05, low CI limit = −0.07, high CI limit = − 0.03, Fig 2c), which may suggest greater positive-size selectivity in the artificially illuminated areas.

**CONCLUSIONS**

Our results revealed that even short-lasting exposure to ALAN may affect the vertical distribution of lake zooplankton and increase fish foraging efficiency, which may affect the size structure of zooplankton populations and the structure of zooplankton communities, thus possibly affecting the functioning of the ecosystem as a whole.

**AUTHORS’ CONTRIBUTIONS**

Conceptualization: J.T., M.Š. and P.M., methodology: J.T., M.Š.; validation: J.T., M.Š. and P.M.; formal analysis, J.T., K.R.; investigation, J.T., M.Š., P.M. and E.B.; writing: J.T., M.Š. and P.M., funding acquisition, J.T. and P.M.

**SUPPLEMENTARY DATA**

Supplementary data is available at *Journal of Plankton Research* online.

**ACKNOWLEDGEMENTS**

We would like to thank (Late.) A. Hankiewicz and W. Pawlak for technical support. The study was conducted with the help of the infrastructure of the Hydrobiological Station of the University of Warsaw in Pilchy.

**FUNDING**

Polish National Science Centre (2016/21/N/NZ8/00914, 2016/23/D/NZ8/03532, 2019/35/B/NZ8/04523).

**REFERENCES**

Becker, A., Whitfield, A. K., Cowley, P. D., Järnegren, J. and Næsje, T. F. (2013) Potential effects of artificial light associated with anthropogenic infrastructure on the abundance and foraging behaviour of estuary-associated fishes. *J. Appl. Ecol.*, 50, 43–50. [https://doi.org/10.1111/j.1365-2664.12024](https://doi.org/10.1111/j.1365-2664.12024).
Bolton, D., Mayer-Pinto, M., Clark, G. F., Dafforn, K. A., Brasil, W. A., Becker, A. and Johnston, E. L. (2017) Coastal urban lighting has ecological consequences for multiple trophic levels under the sea. *Sci. Total Environ.*, 576, 1–9. https://doi.org/10.1016/j.scitotenv.2016.10.037.

Czarnecka, M., Kakareko, T., Jermacz, Ł., Pawlak, R. and Kobak, J. (2019) Combined effects of nocturnal exposure to artificial light and habitat complexity on fish foraging. *Sci. Total Environ.*, 684, 14–22. https://doi.org/10.1016/j.scitotenv.2019.05.280.

Fermin, A. C. and Seronay, G. A. (1997) Effects of different illumination levels on zooplankton abundance, feeding periodicity, growth and survival of the Asian seabass, *Lates calcarifer* (Bloch), fry in illuminated floating nursery cages. *Aquaculture*, 157, 227–237. https://doi.org/10.1016/S0044-8486(97)00167-1.

Gaston, K. J., Visser, M. E. and Hölker, F. (2015) The biological impacts of artificial light at night: the research challenge. *Philos. Trans. R. Soc. Lond. Ser. B Biol. Sci.*, 370, 20140133. https://doi.org/10.1098/rstb.2014.0133.

Gliwicz, Z. M. (2003) Between hazards of starvation and risk of predation: the ecology of offshore animals. In Kinne, O. (ed.), *Excellence of Ecology, Book 12*, International Ecology Institute, Oldendorf/Luhe.

Longcore, T. and Rich, C. (2004) Ecological light pollution. *Front. Ecol. Environ.*, 2, 191–198. https://doi.org/10.1890/1540-9295(2004)002[0191:ELP]2.0.CO;2.

Ludvigsen, M., Berge, J., Geoffroy, M., Cohen, J. H., de la Torre, P. R., Nornes, S. M., Singh, H., Sørensen, A. J. et al. (2018) Use of an autonomous surface vehicle reveals small-scale diel vertical migrations of zooplankton and susceptibility to light pollution under low solar irradiance. *Sci. Adv.*, 4, eaap9887. https://doi.org/10.1126/sciadv.aap9887.

Maszczyk, P., Tałanda, J., Babkiewicz, E., Leniowski, K. and Urban, P. (2021) Daphnia depth selection in gradients of light intensity from different artificial sources: an evolutionary trap? *Limnol. Oceanogr.*, 66, 1367–1380. https://doi.org/10.1002/lno.11691.

Moore, M. V., Pierce, S. M., Wahh, H. M., Kvalvik, S. K. and Lim, J. D. (2000) Urban light pollution alters the diel vertical migration of *daphnia*. *Vet. Internat. Verin. Thes. Angew. Limnol.*, 27, 779–782. https://doi.org/10.1080/03680770.1998.11901341.

Nguyen, K. Q. and Winger, P. D. (2019) Artificial light in commercial industrialized fishing applications: a review. *Rev. Fish. Sci. Aquac.*, 27, 106–126. https://doi.org/10.1080/23308249.2018.1496065.

Nuñez, J. D., Bas, C. C., Perez Garcia, M., Ocampo, E. H., Ribeiro, P. D. and Luppi, T. A. (2021) Artificial light at night may increase the predation pressure in a salt marsh keystone species. *Mar. Environ. Res.*, 167, 105285. https://doi.org/10.1016/j.marenvres.2021.105285.

Owens, A. C. and Lewis, S. M. (2018) The impact of artificial light at night on nocturnal insects: a review and synthesis. *Ecol. Evol.*, 8, 11337–11358. https://doi.org/10.1002/ece3.4557.

Rudstam, L. G., Melnick, N. G., Timoshkin, O. A., Hansson, S., Pushkin, S. V. and Nemov, V. (1992) Diet dynamics of an aggregation of *Macrostemus bramichii* (Dyble) (Amphipoda, Gammaridae) in the Bargusin Bay, Lake Baikal, Russia. *J. Great Lakes Res.*, 18, 286–297. https://doi.org/10.1016/S0380-1330(92)71296-9.

Sanders, D., Frago, E., Kehoe, R., Patterson, C. and Gaston, K. J. (2021) A meta-analysis of biological impacts of artificial light at night. *Nat. Ecol. Evol.*, 5, 74–81. https://doi.org/10.1038/s41559-020-01322-x.

Sanders, D. and Gaston, K. J. (2018) How ecological communities respond to artificial light at night. *J. Exp. Zool. A Ecol. Integr. Physiol.*, 329, 394–400. https://doi.org/10.1002/jez.2157.

Springer, A. and Skrzypczak, A. (2015) The effect of above–water artificial light on the zooplankton abundance in cages for fish rearing. *Pol. J. Nat. Sci.*, 30, 379–387.

Szczerbowski, J. A. and Mamicz, A. (1984) Rearing of coregonid fishes (Coregonidae) in illuminated lake cages: II. Environmental conditions during fish rearing. *Aquaculture*, 40, 147–161. https://doi.org/10.1016/0044-8486(84)90552-1.

Tałanda, J., Maszczyk, P. and Babkiewicz, E. (2018) The reaction distance of a planktivorous fish (*Sardinia erythrophthalmus*) and the evasiveness of its prey (*Daphnia pulicaria pulicaria*) under different artificial light spectra. *Limnology*, 19, 311–319. https://doi.org/10.1007/s10201-018-0548-0.

Zapata, M. J., Sullivan, S. M. P. and Gray, S. M. (2019) Artificial lighting at night in estuaries—implications from individuals to ecosystems. *Estuar. Coast.*, 42, 309–330. https://doi.org/10.1007/s12237-018-0479-3.