Construction of Effective Minimal Active Microbial Consortia for Lignocellulose Degradation

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Abstract
Enriched microbial communities, obtained from environmental samples through selective processes, can effectively contribute to lignocellulose degradation. Unfortunately, fully controlled industrial degradation processes are difficult to reach given the intrinsically dynamic nature and complexity of the microbial communities, composed of a large number of culturable and unculturable species. The use of less complex but equally effective microbial consortia could improve their applications by allowing for more controlled industrial processes. Here, we combined ecological theory and enrichment principles to develop an effective lignocellulose-degrading minimal active microbial Consortia (MAMC). Following an enrichment of soil bacteria capable of degrading lignocellulose material from sugarcane origin, we applied a reductive-screening approach based on molecular phenotyping, identification, and metabolic characterization to obtain a selection of 18 lignocellulose-degrading strains representing four metabolic functional groups. We then generated 65 compositional replicates of MAMC containing five species each, which vary in the number of functional groups, metabolic potential, and degradation capacity. The characterization of the MAMC according to their degradation capacities and functional diversity measurements revealed that functional diversity positively correlated with the degradation of the most complex lignocellulosic fraction (lignin), indicating the importance of metabolic complementarity, whereas cellulose and hemicellulose degradation were either negatively or not affected by functional diversity. The screening method described here successfully led to the selection of effective MAMC, whose degradation potential reached up 96.5% of the degradation rates when all 18 species were present. A total of seven assembled synthetic communities were identified as the most effective MAMC. A consortium containing Stenotrophomonas maltophilia, Paenibacillus sp., Microbacterium sp., Chryseobacterium taiwanense, and Brevundimonas sp. was found to be the most effective degrading synthetic community.

Keywords Lignocellulose · Degradation · Microbial consortium · Functional diversity

Introduction
The biological degradation of lignocellulosic waste materials for subsequent energy production is considered a very promising and sustainable way to supply energy demands. For instance, lignocellulose agrowaste, such as straw and bagasse from sugarcane production, can potentially be used in industry, generating a wide range of value-added bioproducts (e.g., biogas, enzymes, antioxidants) and biofuels [1]. This degradation process relies on the breakdown of lignocellulose—composed of hemicelluloses, cellulose, and the recalcitrant aromatic compound, lignin—through chemical, enzymatic, or thermomechanical processes that convert the polysaccharides into their constituent sugars [2].
The use of microbial consortia is considered an effective and sustainable way of promoting lignocellulose degradation, demonstrating enhanced degradation potential when compared to monoculture approaches, i.e., individual isolates [3–5]. Microbial consortia can be defined as communities ranging from two species communities to undefined, multispecies aggregations, where microorganisms use diverse mechanisms involving multiple complementary enzymes, particularly glycoside hydrolases (GHs), to deconstruct hemicellulose and cellulose. Lignin depolymerization on the other hand is achieved by using peroxidases and laccases [6]. Thus, the degradation of lignocellulose by microbial communities involves several complex and sometimes overlapping mechanisms, which are in general the result of a common effort by the members of the microbial community, due to resource complementarity.

Relationships among species diversity, stability, and function have been central topics in microbial ecology for several decades. Mounting evidence has demonstrated that the direct and positive relationship between function and diversity [7, 8] is often driven by the high complementarity between species, where higher productivity is observed in functionally diverse communities, as determined by their metabolic potential [9, 10]. In addition, ecological evidence supports the contention that microorganisms in a community depend on the activity of other microorganisms to grow, adapt, and reproduce [11–13]. To do so, they make up of complementary mechanisms involving the acquisition and exchange of metabolites to efficiently extract available energy resources. Thus, mixed populations are expected to perform functions that are difficult or even impossible for single species [14] while dealing with potential environmental fluctuations. These ecological principles underlie the higher effectiveness of lignocellulose-degrading microbial consortia and are currently being used to engineer microbial consortia able to effectively perform a range of processes [15].

The enrichment culture technique is a powerful tool to obtain microbial consortia with desired degradation properties [16, 17] because microbial consortia obtained with this method are closer to those functioning in nature [16]. Thus, in the last decades, several lignocellulose-degrading microbial consortia enriched from environmental samples have been identified and functionally characterized [18–22], revealing that functional redundancy is present in these systems [23]. Up to now, it is clear that microbial consortia production systems must account for the environmental relationships, distribution, abundance, and functional diversity of the participating members and their specific role during degradation [24].

Even though the analysis of enriched communities from environmental samples reveals the dominance of few bacterial phyla (e.g., Proteobacteria, Firmicutes, Chloroflexi, and Bacteroidetes), the number of ecological dominant species (culturable and unculturable) remains high, i.e., > 20 identified strains [18, 25–29]. The complexity of such systems can make it difficult to disentangle the interactive network that ultimately drives the degradation process. Importantly, the lack of knowledge on the interactions between these large numbers of species during degradation hampers the upscaling of the consortia to an industrial system designed to obtain lignocellulose biodegradation and derived products like biofuels. In addition, the fact that not all effective strains from the enrichments can be isolated and grown in the lab limits their application for industrial purposes. Thus, designing an effective degrading consortium harboring a reduced number of culturable members could be advantageous, enabling a full characterization of the structural composition, the degradative power, and the interactive roles of degradation players. In other words, a better understanding of the degradation process could pave the way for a more efficient breakdown of lignocellulose for environmental and commercial purposes.

In order to obtain a minimal active microbial consortium (MAMC) from environmental samples, we used an ecologically driven, reductive-screening approach (reducing the number of species throughout the investigation), starting with the isolation of lignocellulose-degrading strains obtained by selection through an enrichment process of soil bacteria grown in sugarcane-biowaste lignocellulose substrates, followed by molecular phenotyping, identification, and metabolic characterization. Based on the metabolic characterization, 45 bacterial strains were classified as belonging to four functional metabolic groups. A set of 18 strain representatives of these groups were further used to construct a total of 65 synthetic communities containing five species each (65 compositional replicate MAMCs). Thus, MAMCs varied in their functional diversity, as determined by the number of functional metabolic groups, as well as in metabolic and degradation potential, while remaining constant in species richness. The MAMCs were evaluated for their degradation capacity, under the hypothesis that MAMC with higher functional diversity would lead to higher degradation rates.

**Materials and Methods**

**Isolation and Maintenance of Strains**

In order to obtain bacteria with high lignocellulose degradation capacity, we performed an enrichment experiment [28–30] using two sugarcane related substrates—bagasse (B) and straw (St)—and a soil inoculum obtained from a sugarcane plantation, which generated a total of 18 flasks including controls with substrate and no inoculum. Briefly, 10 g of soil inoculum was used to prepare a soil suspension by adding to 90 mL of sodium chloride 0.90% and 10 g of sterile gravel in 250-mL flasks. After shaking for 1 h at 250 rpm at room
temperature (20 °C), aliquots of 250μL were inoculated to triplicate 100-mL flasks containing 25 mL of mineral salt sterile medium (MSM) [30] with 1% of the lignocellulose sterile substrate. Cell density was controlled microscopically at regular time intervals and when cultures reached 10^9 cells/mL, an aliquot of 25 μL of culture was transferred into 25 mL of fresh medium. After ten transfers and after the final flasks reached 10^9 cells/mL, the enrichment process was stopped. A 2-mL sample from several flasks of lignocellulose-enriched soil community’s final transfer (T10) was then used for colony isolation (for the enrichment experiment, see supplementary Fig. S1(a)). A homogenized sub-sample of 1 mL was used for serial dilutions in NaCl 0.85% until 10^5 and plated onto minimal nutrient medium agar plates (Reasoner’s 2A agar, R2A) (Becton Dickenson, Cockeysville, MD) in triplicates. Colonies were chosen according to morphological uniqueness via visual inspection and were further purified using several transfer steps and maintained in R2A agar. For long-term preservation and further studies, fresh biomass of the isolates obtained with LB broth (72 h of growth) was suspended in 20% glycerol and stored at −20 °C.

Genotypic Differentiation Using BOX-PCR (Molecular Fingerprinting)

We performed colony BOX-PCR in order to compare the genomic profiles of all isolated colonies, and select for individual strains. BOX-PCR was performed with the primer A1R (CTACGGCAGGCGACGCTGA) and the following PCR amplification conditions: 95 °C for 2 min; 35 cycles of 94 °C for 30s, 50 °C for 1 min, and 65 °C for 8 min; and a final extension step at 65 °C for 16 min [31]. PCR products were then run in 1.5% agarose gel and further analyzed (cluster analysis of BOX-PCR pattern) in GelCompar II (Applied Maths) using Dice coefficient and the UPGMA clustering. Isolates with a >95% homology were assigned as belonging to the same strain.

Phylogenetic Identification of Isolates by Partial 16S rRNA Gene Sequencing

The DNA extraction from purified cultures of a total of 96 representative strains that belonged to unique BOX groups was performed with the UltraClean® Microbial DNA Isolation Kit (MoBio® Laboratories Inc., Carlsbad, CA, USA) according to the manufacturer’s instructions. We assigned an ID number to each strain (ID strain number 1 to 96) for identification purposes during the study. Bacterial 16S rRNA genes were amplified using primers B8F (5′-AGAGTTTGATCMTGGCTCAG-3′ [32]) and U1406R (5′-ACGCGGCGGTGTGTRC-3′ [33]). PCR reactions were done in a 50-μL reaction mixture following the protocol of Taketani et al. [34]. PCR products were sequenced by Sanger technology (LGC Genomics, Germany). All resulting chromatograms were analyzed and trimmed for quality using the Lucy algorithm (http://rdp.cme.msu.edu/). Taxonomic assignment of the sequences (>99% identity) was done using BLAST-N against the National Center for Biotechnology Information (NCBI) database (http://blast.ncbi.nlm.nih.gov/Blast.cgi) and confirmed at RDP (Ribosomal database project).

Enzymatic Activities Related to Hemicellulose and Cellulose Degradation

In order to determine distinct enzymatic degradation potentials of the strains identified by BOX-PCR and 16S rRNA gene sequencing, we quantified the enzymatic activities related to hemicellulose and cellulose breakdown during growth in MSM media with 1% of the specific lignocellulosic substrate from which they were originally obtained. Microbial cells plus substrate from 2-mL samples were harvested by centrifugation for 5 min at 10000 rpm after 72 h of growth. The supernatant (secretome) was recovered and tested for enzymatic activity in triplicates using MUF-β-D-xylpyranoside, MUF-β-D-mannopyranoside, MUF-β-D-galactopyranoside, MUF-β-D-celllobioside, and MUF-β-D-glucopyranoside as substrates. For the reaction, 10 μL of MUF-substrate (10 mM in dimethyl sulfoxide), 15 μL of Melleivane buffer (pH 6.8), and 25 μL of each supernatant were mixed and incubated at 28 °C for 45 min in the dark. The reaction was stopped by adding 150 μL of 0.2 M glycine-NaOH buffer (pH 10.4). Fluorescence was measured at an excitation of 365 nm and emission of 445 nm. The negative controls consisted in sterile PCR water and a mixture without the MUF substrates. Enzyme activities were determined from the fluorescence units using a standard calibration curve built with glycine buffer (pH 10.4) and expressed as rates of MUF production (μM MUF per min at 28 °C pH 6.8).

Phenotypic Diversity (Biolog GEN III Plates)

The metabolic profiles of 45 enzymatically active selected strains were monitored during growth on 71 energy sources of GEN III Biolog plates (Biolog Inc., Hayward, CA). The absorbance values at 590 nm were measured using a microplate reader (VersaMax microplate reader; Molecular Devices Corp.). For each strain, colonies obtained on R2A agar were pulled into inoculation fluid (IF-B) at an optical density (OD590) of 0.03. After 2 h of starvation in the IF-B fluid, 100 μL was transferred to each well of the GEN III plates including one blank. The plates were incubated at 28 °C and read every 12 h until 72 h of growth. The normalized sum of all measured time points for the well was used to calculate the area under the curve for each carbon source. These data were used to generate a principal component analysis (PCA) using...
Canoco v5.0 [35], which allowed the classification of strains into functional metabolic groups (FMG), by calculating the similarities in metabolic profiling across all 71 substrates.

**MAMC**

A total of 18 strains representing all FMG were used to create compositional replicates of five-species MAMC. Specifically, a total of 65 different synthetic communities were created by combining members of two, three, or all four FMG (Table 1), thus generating MAMC with different levels of functional diversity (two, three, or four FMG) but similar levels of species richness. For each level of functional diversity, we prepared three to four compositional replicates of each synthetic community, i.e., using the same number of species from the same FMG but different species, leading a total of 65 synthetic communities. The use of compositional replicates allowed for proper quantification of functional diversity on the substrate’s degradation while controlling for the influence of species identity.

After selecting the composition of each one of the 65 MAMCs, these were constructed by mixing all strains in equal concentrations. Briefly, we grew the strains in R2A broth during 40 h (180 rpm, 28 °C) and adjusted each strain to an OD of 0.02 in NaCl 0.85%. An aliquot of 125 μL of each diluted strain (approximately 1 × 10^5 cells/mL, as determined by plate counting) was inoculated into 25 mL of MSM medium with 1% of straw (total cell concentration of 5 × 10^5 cells/mL) and samples was defined with a resolution of 4 cm⁻¹ under ambient atmosphere at room temperature and in triplicates. The spectra were integrated and baseline corrected using the Spectrum™ software. The analysis was performed using the Unscrambler X (Camo Software Oslo Norway). A 5-point Savitzky–Golay smoothing algorithm was applied to the calibration set’s spectra and used to predict concentrations in the samples (using partial least squares (PLS) regression). The predicted composition of each sample obtained with PLS was expressed as the percentage of degradation (from the initial amount; %D) and was calculated for each lignocellulose fraction of the lignocellulose as follows: %D = [(a – b) / a] × 100; where a = percentage of the fraction in the substrate before incubation; b = percentage of the fraction in the substrate after incubation.

**Functional Diversity Measurements**

The community niche (CN) of a given MAMC was obtained based on the performance of each species on each one of the 71 carbon sources from the Biolog GEN III plate [9]. The CN value corresponds to the sum of the best performances per carbon source found in each synthetic community. Furthermore, we used the metabolic potential of each strain, also based on Biolog data, to calculate the functional attribute distance (FAD), used as a proxy of functional diversity [37, 38]. FAD was calculated by using Euclidean distance to calculate the pairwise distance between species (Function dist()) implemented in RStudio 1.0.136 [39].

**Results**

**Isolation and Identification of Enriched Bacterial Strains**

Samples from the end-point populations (Transfer 10) of all the cultures with substrate (including the controls) were diluted and plated onto R2A agar. A total of 157 isolates (up to 16 colonies per flask) were obtained from the last three plated dilutions. Molecular characterization of the purified strains using BOX PCR revealed a total of 96 unique groups across all samples, which were identified by sequencing of the 16S rRNA gene in 72 bacterial species (see Table S1 supplementary information for identity and GeneBank accession numbers).

**Enzymatic Degradation Potentials**

The enzymatic degradation potential of 72 strains was measured after 72 h of growth under the enriched correspondent substrate. Results did not have a correlation with the substrate type in all MUF substrates (p > 0.05). The highest enzymatic activity in all secretomes was observed in the β-D-mannopyranoside activity (related to hemicellulolytic degradation) with an averaged 1.17 μM MUF/min followed by 0.2 μM MUF/min of β-D-celllobioside (involved in cellulose degradation) and 0.12 μM MUF/min of β-D-xylopyranoside (as part of hemicellulolitic activities). Based on enzymatic
Table 1  Description of the MAMC according to the number of functional metabolic groups. Type of mix according to the number of functional metabolic groups (FMG) combined (FMG I, FMG II, FMG III, FMG IV) and ID number of strains used in each synthetic community per functional group. All synthetic communities were at least three compositional replicates. m: identification of the synthetic community. For strain identification, see Table 2

| Type of Mix | Synthetic community | FMG - Strains ID number |
|-------------|---------------------|-------------------------|
|             |                     | I | II | III | IV |
| All 18 strains |                     |   |    |     |    |
| Type A: Mixing four FMG |                     |   |    |     |    |
|                     | m1                  |   |    |     |    |
|                     | m2                  | 76 | 10 | 49  | 68 | 41 |
|                     | m3                  | 76 | 26 | 90  | 11 | 44 |
|                     | m4                  | 76 | 62 | 57  | 29 | 25 |
|                     | m5                  | 10 | 62 | 85  | 52 | 15 |
|                     | m6                  | 10 | 49 | 90  | 11 | 25 |
|                     | m7                  | 62 | 49 | 57  | 52 | 44 |
|                     | m8                  | 76 | 90 | 85  | 18 | 41 |
|                     | m9                  | 10 | 49 | 85  | 29 | 13 |
|                     | m10                 | 76 | 49 | 68  | 11 | 41 |
|                     | m11                 | 10 | 90 | 68  | 29 | 25 |
|                     | m12                 | 26 | 57 | 68  | 52 | 44 |
|                     | m13                 | 62 | 85 | 29  | 18 | 15 |
|                     | m14                 | 10 | 85 | 68  | 41 | 44 |
|                     | m15                 | 26 | 90 | 11  | 41 | 25 |
|                     | m16                 | 76 | 49 | 29  | 44 | 15 |
|                     | m17                 | 62 | 57 | 52  | 13 | 41 |
| Type B: mixing two FMG |                     |   |    |     |    |
|                     | m18                 | 76 | 10 | 49  | 57 | 13 |
|                     | m19                 | 76 | 10 | 49  | 85 | 13 |
|                     | m20                 | 76 | 10 | 57  | 85 | 13 |
|                     | m21                 | 76 | 10 | 49  | 57 | 85 |
|                     | m22                 | 76 | 10 | 49  | 57 | 85 |
|                     | m23                 | 10 | 62 | 49  | 57 | 85 |
|                     | m24                 | 76 | 10 | 49  | 57 | 85 |
|                     | m25                 | 76 | 10 | 49  | 57 | 85 |
|                     | m26                 | 76 | 10 | 49  | 57 | 85 |
|                     | m27                 | 76 | 10 | 49  | 57 | 85 |
|                     | m28                 | 76 | 10 | 49  | 57 | 85 |
|                     | m29                 | 10 | 62 | 49  | 57 | 85 |
|                     | m30                 | 76 | 62 | 68  | 18 | 13 |
|                     | m31                 | 76 | 62 | 68  | 29 | 13 |
|                     | m32                 | 76 | 62 | 68  | 29 | 13 |
|                     | m33                 | 76 | 62 | 68  | 29 | 13 |
|                     | m34                 | 76 | 62 | 68  | 29 | 13 |
|                     | m35                 | 10 | 62 | 68  | 29 | 13 |
|                     | m36                 | 76 | 10 | 68  | 18 | 13 |
|                     | m37                 | 76 | 10 | 68  | 29 | 13 |
|                     | m38                 | 76 | 10 | 68  | 29 | 13 |
|                     | m39                 | 76 | 10 | 68  | 29 | 13 |
|                     | m40                 | 76 | 10 | 68  | 29 | 13 |
|                     | m41                 | 76 | 10 | 68  | 29 | 13 |
|                     | m42                 | 76 | 10 | 68  | 29 | 13 |
|                     | m43                 | 76 | 10 | 68  | 29 | 13 |
|                     | m44                 | 76 | 10 | 68  | 29 | 13 |
|                     | m45                 | 76 | 10 | 68  | 29 | 13 |
|                     | m46                 | 76 | 10 | 68  | 29 | 13 |
|                     | m47                 | 76 | 10 | 68  | 29 | 13 |
|                     | m48                 | 76 | 10 | 68  | 29 | 13 |
|                     | m49                 | 76 | 10 | 68  | 29 | 13 |
|                     | m50                 | 76 | 10 | 68  | 29 | 13 |
|                     | m51                 | 76 | 10 | 68  | 29 | 13 |
|                     | m52                 | 76 | 10 | 68  | 29 | 13 |
|                     | m53                 | 76 | 10 | 68  | 29 | 13 |
| Type C: Mixing 3 FMG |                     |   |    |     |    |
|                     | m54                 | 76 | 10 | 68  | 29 | 13 |
|                     | m55                 | 76 | 10 | 68  | 29 | 13 |
|                     | m56                 | 76 | 10 | 68  | 29 | 13 |
|                     | m57                 | 76 | 10 | 68  | 29 | 13 |
|                     | m58                 | 76 | 10 | 68  | 29 | 13 |
|                     | m59                 | 76 | 10 | 68  | 29 | 13 |
|                     | m60                 | 76 | 10 | 68  | 29 | 13 |
|                     | m61                 | 76 | 10 | 68  | 29 | 13 |
|                     | m62                 | 76 | 10 | 68  | 29 | 13 |
|                     | m63                 | 76 | 10 | 68  | 29 | 13 |
|                     | m64                 | 76 | 10 | 68  | 29 | 13 |
|                     | m65                 | 76 | 10 | 68  | 29 | 13 |
activity potential, we selected a total of 25 strains from bagasse and 20 strains from straw (45 strains) (Fig. S2 Supplementary information; identity of these 45 strains can be found in Table S1).

**Metabolic Diversity and Construction of Consortia**

In order to construct the MAMC, we first characterized the metabolic profile of the 45 selected strains using Biolog GEN III plates. A principal component analysis (PCA) performed with the data obtained after 72 h of incubation revealed distinct metabolic profiles that were not associated with the substrate type used for the enrichment (bagasse or straw) (Fig. 1). Most of the variation is explained by the first axis (35.66%); however, variation is moderately explained by the second axis (17.36%). We clustered the PCA results according to the four resulting sections in the plot (I, II, III, and IV denoted by different colors in Fig. 1), thus determining four distinct FMG. We noticed that according to the nature of the Biolog substrates (Table S2 supplementary information), functional metabolic groups III and IV have a preference for carbohydrate metabolism. Groups I and II, on the other hand, have a preference for amino acids and carboxylic acids.

A total of 18 representative strains from each of the FMG (see Table 2) were then used to construct five-species MAMC, thus creating a total of 65 synthetic communities by combining members of two, three, or all four functional metabolic groups (Table 1).

**Degradation of Lignocellulose Fractions and Correlation with Metabolic Potentials**

FTIR analysis performed on dried substrate obtained after consortia’s growth depicts degradation of lignocellulose per lignocellulose’s fraction (lignin, cellulose, and hemicellulose). Figure 2 shows the percentage of degradation of all 65 synthetic communities including three replicates containing all 18 strains per lignocellulosic fraction. From the results, we could distinguish that the synthetic communities type A (MAMC with four functional groups, m2-m17) degraded lignin effectively when compared to the degradation of hemicellulose and cellulose. On the other hand, type C synthetic communities (MAMC with three functional groups, m54-m65) have a better degradation of hemicellulose. A MDS (Fig. S3 Supplementary information) plotting mixes type A and type C shows a fair separation between these two types. Synthetic communities type B (MAMC with two functional groups, m18-m53), however, do not have a clear clustering since the degradation levels of all three lignocellulosic fractions were rather heterogeneous among communities (data not plotted). The synthetic community containing all 18 strains shows the highest degradation profile of all lignocellulosic fractions (52% on average).

In order to measure the degree of functionality across the synthetic communities, and their potential effect on lignocellulose degradation, we used three FD measures: (i) FMG, which represents the number of functional groups in the synthetic community; (ii) CN; and (iii) FAD, both of which are associated with the metabolic potential of the community. Our results revealed that both the FD and the lignocellulose component influenced the relationship (Fig. 3). FMG did not generate any significant pattern, although a positive relationship between lignin degradation and FMG was close to significant ($p$ value = 0.061), a positive tendency was also observed in CN results. A positive relationship was equally observed between the degradation of lignin ($p$ value = 0.025) and FAD, although the explanatory power was very low ($R^2 = 0.0238$). Conversely, higher metabolic diversity (CN) had a negative and significant effect on the degradation of cellulose and negative tendency on the degradation of hemicellulose, showing here a similar pattern than those obtained with FMG and FAD.
Identification of Potential Strains and the Best MAMC

We identified a total of seven synthetic communities with an averaged degradation > 30% (maximum value 50.6%) in all three lignocellulosic fractions: m5, m22, m28, m33, m48, m53, and m61. A RDA triplot using the degradation results data vs the combination of strains (as environmental data) revealed that the seven synthetic communities with relatively high degradation clustered together and shared specific strains (Fig. 4). Table 3 summarizes the degradation results of the seven synthetic communities.

The most effective lignocellulose-degrading strains common among the seven selected synthetic communities were Stenotrophomonas maltophilia strain JN40 (ID Number 13), Paenibacillus sp. PALXIL05 (ID Number 18), Microbacterium sp. UYFA68 (ID Number 25), Chryseobacterium taiwanense strain DUCC3723 (ID Number 29), and Cellulosimicrobium sp. BAB-2381 (ID Number 76).

Table 2

| Functional metabolic group | ID strain no. | Identification                                      |
|----------------------------|--------------|-----------------------------------------------------|
| I                          | 76           | Cupriavidus pauculus partial 16S rRNA gene strain KPS201 |
|                            | 10           | Pseudomonas plecoglossicida strain ICB-TAB78 16S rRNA gene |
|                            | 26           | Alcaligenes sp. DA5 16S ribosomal RNA gene          |
|                            | 62           | Paracoccus sp. B160 16S ribosomal RNA gene          |
| II                         | 49           | Achromobacter sp. HBCCD-1 16S ribosomal RNA gene   |
|                            | 90           | Devosia riboflavina strain HPG62 16S ribosomal RNA gene |
|                            | 57           | Ochrobactrum sp. 71B2 16S ribosomal RNA gene        |
|                            | 85           | Sphingobacterium sp. Bt-34 16S ribosomal RNA gene  |
| III                        | 68           | Brevundimonas sp. R3 16S ribosomal RNA gene         |
|                            | 11           | Cellulosimicrobium sp. BAB-2381 16S ribosomal RNA gene |
|                            | 29           | Chryseobacterium taiwanense strain DUCC3723 16S ribosomal RNA gene |
|                            | 52           | Flavobacterium sp. WG1 partial 16S rRNA gene strain WG1 |
|                            | 18           | Paenibacillus sp. PALXIL05 16S ribosomal RNA gene  |
| IV                         | 41           | Enterobacter aerogenes strain K_G_AN-5 16S ribosomal RNA gene |
|                            | 44           | Pseudomonas sp. GT 2-02 16S ribosomal RNA gene      |
|                            | 25           | Microbacterium sp. UYFA68 16S ribosomal RNA gene    |
|                            | 15           | Bacillus nealsonii strain RT55b2 16S ribosomal RNA gene |
|                            | 13           | Stenotrophomonas maltophilia strain JN40 16S ribosomal RNA gene |

Fig. 2 Percentage of degradation (%D) of the three lignocellulosic fractions obtained with FTIR for each of the MAMC—synthetic communities of five species (m2-m65), taken from a pool of 18 selected lignocellulose degrading strains. The first three bars (1a–1c) represent the degradation potential of all 18 strains combined.
strain DUCC3723 (ID Number 29), Paracoccus sp. B160 (ID Number 62), Brevundimonas sp. B1 (ID Number 68), Burkholderia sp. YX02 16S ribosomal RNA gene (ID Number 76). According to the degradation results, the highest degradation was observed in the synthetic community number 48 (degradation averaged of all lignocellulose fractions 50.6%) containing strains with ID numbers 13, 18, 25, 29, and 68. All the strains in this synthetic community belonged to the functional metabolic groups with preference for carbohydrate metabolism (groups III and IV).

### Table 3

| Synthetic community | Lignin   | Cellulose | Hemicellulose |
|---------------------|----------|-----------|---------------|
| m5                  | 44.21    | 30.74     | 26.93         |
| m22                 | 40.79    | 25.51     | 45.09         |
| m28                 | 25.08    | 36.73     | 32.28         |
| m33                 | 37.56    | 36.84     | 37.05         |
| m48                 | 43.01    | 39.65     | 69.18         |
| m53                 | 42.76    | 37.83     | 32.38         |
| m61                 | 40.32    | 36.67     | 25.81         |
Discussion

In this study, we used a reductive-screening approach coupled with ecological strategies to obtain minimal active microbial consortia (MAMC) capable of effectively degrading lignocellulose, using bacterial strains obtained from environmental (soil) samples. Our reductive method included an enrichment experiment and the isolation of final community members. Furthermore, we used enzymatic and metabolic-profile assessments to determine the degradation potential and metabolic property of the isolated members of the community. Enzymatic assessments are rather simple, powerful, and quantitative approaches for screening purposes and have been effectively used in analyzing lignocellulose degradation capacities [26, 29, 40]. Here, we used this method for screening purposes, focusing only on enzymes from the GH family, which are related to hemicellulose and cellulose degradation. Furthermore, the use of metabolic profiling tools like Biolog was advantageous in depicting metabolic preferences of the initially selected strains, define functional metabolic groups, identify the metabolic preferences of the effective degrading bacteria, and reveal the relationship between functional diversity and degradation.

Using the metabolic and enzymatic characterization of the strains together with 16S rRNA gene sequencing, we further designed 65 MAMC (bacterial communities containing synthetically diverse communities of only microbial species) that were functionally diverse, i.e., displaying various levels of complementarity in terms of both carbon source utilization and enzymatic activities associated with the different lignocellulose components. These functionally complementary communities were then assessed for their capacity to degrade all three lignocellulose fractions (see Fig. S1(b) in the supplementary information for a schematic representation of the reductive screening approach). Our aim was to determine whether high functional diversity would lead to higher degradation potential, as predicted in community ecology theory [41], given that high functionality is expected to lead to higher number of (complementary) functions or ecological roles needed for an effective degradation.

Our results showed that the number of strains included in the consortia did not determine the effectiveness of degradation since the results obtained with all 18 strains were comparable to the results obtained with several five-strain constructed synthetic communities (the synthetic community with the maximum averaged degradation reached up to 96.5% of the degradation rate when compared to the synthetic community containing all 18 strains). Other studies obtained similar degradation levels (between 40 and 50% average of all three lignocellulose fractions [28, 29] with more than 20 different identified strains in the community). Here, we could identify a total of seven synthetic communities with optimal degradation levels. Results were mostly identity dependent, but there were positive and negative connections with functional diversity. We can then conclude that reducing the number of community members provides a better overview of functional diversity and ecological roles without affecting the degradation levels. In another notice, we found that higher functional diversity increased the degradation of the most complex substrate (lignin). Lignin’s breakdown involves the release of a great variety of compounds including complex aromatic carboxylic acids, which might eventually converge into the citric acid cycle [42]. Our results suggest that the strains with the enzymatic machinery able to metabolize lignin metabolic products (aromatic carboxylic acids) might also have the enzymatic means to metabolize complex substrates like lignin.

Thus, the so-called division of labor together with diversity in our built consortia might be acting beneficially upon the potential of the community during lignin degradation. Our results showed that a higher functional diversity could expand and add complementarity for resource use among taxa in the context of bacterial community functioning [38]. On the other hand, the degradation levels of cellulose negatively correlated with the diversity measures based on overall carbon metabolism (FAD and CN) whereas non-significant but negative trends were observed for hemicellulose. The enzymatic degradation of these substrates is accomplished via the collective action of multiple carbohydrate-active enzymes, typically acting together as a cocktail with complementary, synergistic activities and modes of action [6]. This negative relationship could suggest the presence of interspecific competition, which can be a stress factor for the community.

Interestingly, we found commonality in the species among the seven selected synthetic communities with the highest degradation levels, which are frequently reported as effective lignocellulose degraders. For example, Paracoccus sp. was found to be dominant after selection in lignin-amended lignocellulosic cultures; this suggested that Paracoccus sp. has an active role in lignin modification and depolymerization [2, 43]. On the other hand, Burkholderia sp. has been found to secrete specific enzymes capable of degrading plant cell wall components like cellulose, hemicellulose, lignin, and xylose [44, 45]. Another study found specific genes coding for catalases and peroxidases, which might contribute to the lignin degrading ability of Burkholderia sp. [45]. Our RDA analysis showed a positive correlation of Burkholderia sp. with the degradation of all three lignocellulosic fractions.

Other genera included in the most effective synthetic communities were Stenotrophomonas maltophilia, Microbacterium sp., Paenibacillus sp., Chryseobacterium taiwanense, and Brevundimonas sp. The genus Stenotrophomonas occurs ubiquitously in nature and species like S. maltophilia are often found in the rhizosphere and inside many different plant species. Interestingly, Galai et al. [46] demonstrated laccase activity in Stenotrophomonas maltophilia, which was able to deconstruct lignin. Hence, Stenotrophomonas sp. has been found as an excellent candidate for biotechnological applications. On the
other hand, although there are very few reports of Microbacterium sp. strains with lignin degradation ability, Microbacterium species were found to be present in the gut of the wood-infesting termites [47]. There are also reports of strains of Microbacterium able to degrade polycyclic aromatic hydrocarbons [48–50].

Several strains of Paenibacillus sp. have been found to efficiently degrade lignocellulosic fractions [51], having effective lignolytic and cellulolytic enzymes [50, 52, 53]. On the other hand, the role of Chryseobacterium sp. in lignocellulose degradation is still unclear [30]; however, several species of Chryseobacterium have been isolated from lignocellulosic substrates and as part of the degradation processes of cellulose [54].

We found here (Fig. 3) a strong correlation of cellulose degradation with the presence of Paenibacillus sp. and Chryseobacterium sp. in the consortia. Moreover, the presence of Brevundimonas sp. (catalase producers) during lignocellulose degradation has been reported in lignin-amended soils [43] and in enrichment cultures of switchgrass and corn stover [26] and also has been found to be related to cellulose degradation [55].

Here, we were able to construct minimal effective consortia from environmental samples with optimal degradation levels as the ones found in consortia built with a higher number of community members. We found that a consortium containing Stenotrophomonas maltophilia, Paenibacillus sp., Microbacterium sp., Chryseobacterium taiwanense, and Brevundimonas sp. is an effective degrading synthetic community. Further work would be needed to understand the interactive relationship and metabolic pathways of these consortium partners. However, the screening method described here demonstrated to be a useful way to obtain minimal lignocellulose-degrading consortia from environmental samples that can be further developed to efficiently promote a sustainable way of lignocellulose breakdown for environmental or commercial purposes.

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Compliance with Ethical Standards

Conflict of Interest The authors declare that they have no conflict of interest.

Ethical Approval This article does not contain any studies with human participants or animals performed by any of the authors.

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