North American channel catfish, *Ictalurus punctatus*: a neglected but potentially invasive freshwater fish species?

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Received: 2 December 2019 / Accepted: 19 January 2021 / Published online: 12 February 2021 © The Author(s) 2021

Abstract The North American channel catfish *Ictalurus punctatus* has been introduced to several locations in Europe but has received little or no scientific study despite its invasive attributes, including prolific reproduction, tolerance to a wide range of conditions, opportunistic feeding, at least partial ‘predator release’, and some evidence of environmental impacts (e.g. disease transmission). To assess the species’ potential invasiveness and the likely risks to native species and ecosystems in Europe, available literature from both North America and Europe was reviewed and used to carry out risk screenings of the species for the risk assessment areas, North and South Italy, using the Aquatic Invasiveness Screening Kit (AS-ISK), which was followed by a more detailed evaluation (for both North America and Europe) of the species’ potential impacts using the Environmental Impact Classification of Alien Taxa (EICAT) assessment protocol. The AS-ISK score indicated that channel catfish is likely to pose a high risk of being invasive in both North and South Italy, with EICAT scores indicating “Major” impacts for both North America and Europe, at high and medium confidence.
levels, respectively. The present results emphasise the urgent need to carry out in-depth studies on introduced populations of this species to understand better its invasive potential so as to inform management decisions on the appropriate control or eradication measures for invaded water bodies.

**Keywords** Alien species · Invasiveness · Screening · Impact · Management · Assessment

**Introduction**

Of the many non-native fish species that have been introduced to the European continent, the North American channel catfish *Ictalurus punctatus* (Rafinesque, 1818) has received relatively little attention (e.g. Ligas 2007; Haubrock et al. 2018a, b). This contrasts the concerted, comprehensive research on other non-native fishes from North American, such as pumpkinseed *Lepomis gibbosus* (Copp and Fox 2007; Fox and Copp 2014), black bullhead *Ameiurus melas* (Copp et al. 2016a) and crucian carp (Tarkan et al. 2016), and from Asia such as the topmouth gudgeon *Pseudorasbora parva* (Gozlan et al. 2010). This scarcity of scientific study on channel catfish is surprising given its commercial importance to aquaculture and as a sport angling species in both its native and introduced ranges of North America (Brown 1942; Appelget and Smith 1951; Boyd et al. 2000; Rezk et al. 2003; Tucker and Hargreaves 2004; Simmons et al. 2006), as well as in Europe (Carpio et al. 2019). Dedicated studies of channel catfish impacts are limited to the experimental study of Adams (2007), though in a field-based study of non-native fish diet, Poe and Rieman (1988, p. 48) stated that channel catfish “may also be an important predator on juvenile salmonids during the spring in the upper reservoir”.

Channel catfish is considered one of the fastest expanding species (Tucker and Hargreaves 2004; Olden and Poff 2005), characterised by attributes suitable for introductions, including high annual survival rates (Blank et al. 2017) and a wide temperature and salinity tolerances (0–38 °C and 0–11 ppt, respectively; Wellborn 1988; see Supplementary Material 1). With an optimum water temperature range between 24 and 30 °C, channel catfish possesses the demonstrated ability to establish self-sustaining populations outside its native range, both when translocated within North America and introduced elsewhere. Vectors for channel catfish introductions and translocations include aquaculture, pet trade and fishery diversification (sometimes referred to as ‘enhancement’). The development of channel catfish aquaculture in Europe, e.g. Germany (Hilge 1980), led to the species being introduced to at least 25 countries (Supplementary Material 2), and the species now ranks amongst the top 27 of farmed species (Savini et al. 2010). An early database (Welcomme 1988) suggested that channel catfish had become widespread. As with many non-native species, accurate data on established channel catfish populations are lacking. However, given the species’ biology and potential impacts on native species and ecosystems (Haubrock et al. 2018a, b), specific legislation to manage non-native freshwater fishes was enacted in England & Wales in 1998 (www.legislation.gov.uk/uksi/1998/2409/made), and ictalurid catfishes were included on the list of non-native species subject to controls for keeping and release (see Table 1 in Copp et al. 2005a). In the EU, channel catfish has been risk assessed with regard to its use in aquaculture (Copp et al. 2016c). Elsewhere, in the absence of similar controls, slow regulatory and management response times can potentially permit the spread through European freshwater ecosystems (Maguire 2004; Crooks 2005), such as was observed with topmouth gudgeon in Europe (Gozlan et al. 2010).

The potential threat of channel catfish to European freshwater ecosystems is especially relevant to southern Europe, and in Italy, which has been reported to have the highest number of introduced fish species (see Table 2 in Copp et al. 2005a; Nocita et al. 2017), channel catfish populations are known to be established in central and northern Italy. This contrasts the distributions throughout Italy of two other North American catfishes, black bullhead *Ameiurus melas* and brown bullhead *Ameiurus nebulosus* (Bianco and Ketmaier 2001). The aim of the present study was to evaluate the potential invasiveness of channel catfish in Italy as well as the level of potential adverse impacts in Europe as a whole relative to North America as a whole. In the present article, the term ‘invasive’ is used according to the definition given in Copp et al. (2005a). The specific objectives were to: (1) review relevant available information on the possible impacts of the species; (2) carry out risk screenings to rank the
North American channel catfish, *Ictalurus punctatus*

Table 1 Results for channel catfish *Ictalurus punctatus* of risk screenings using the freshwater Fish Invasiveness Screening Kit (FISK) for risk assessment areas in European countries, with risk scores and outcome classifications of: M = medium; MH = moderately high; H = high; VH = very high. Also given are the selection criteria (Crit) used in the cited study and the a priori status (Invasive/Protection), and the associated confidence factor (CF) details.

| Region/country | Crit | a priori status | Score | Outcome | CF | Refs. |
|----------------|------|-----------------|-------|---------|----|-------|
| Greece         | 2    | Invasive/not evaluated | 22.3  | MH      | 0.83 | Perdikaris et al. (2016) |
| Bulgaria       | na   | Invasive/ not evaluated | 10.0  | MH      | 0.79 | Simonovic et al. (2013) |
| England & Wales| na   | Absent/ not evaluated | 23.8  | H       |      | Copp et al. (2009a, b)  |
| Southern Finland| na  | Invasive/ not evaluated | 11.0  | M       | 0.79 | Puntila et al. (2013) |
| Iberian Peninsula | 2   | Invasive/ not evaluated | 31.0  | VH      | 0.81 | Almeida et al. (2013) |
| Lake Balaton, Hungary | 2   | Invasive/least concern | 8.67  | M       | 0.78 | Ferincz et al. (2016) |
| North Italy    | na   | Invasive/ not evaluated | 31.0  | H       | 0.89 | This study |
| South Italy    | na   | Invasive/ not evaluated | 30.0  | H       | 0.92 | This study |

Table 2 Results for channel catfish *Ictalurus punctatus* risk screenings made for Turkey using the freshwater Fish Invasiveness Screening Kit (FISK) and the Aquatic Species Invasiveness Screening Kit (AS-ISK) and for Italy using the AS-ISK, including the status criteria (AQN = Non-native species important for the aquarium trade; NNP = Non-native species already present in the country), the a priori classification (see “Methods”), and the risk screening outcome scores and risk ranks from the Basic Risk Assessment (BRA) and the Climate Change Assessment (CCA). The AS-ISK outcome risk ranks for Turkey are based on the threshold scores (for distinguishing Medium and High risk species) reported in Tarkan et al. (2017) and those used for Italy are given in the “Methods” section.

| Region/country | Criteria | a priori status | AS-ISK | Outcome | CF | Refs. |
|----------------|----------|-----------------|--------|---------|----|-------|
| Turkey         | AQN      | Invasive/least concern | 31.0  | H       | 0.68 | Tarkan et al. (2017) |
| North Italy    | NNP      | Invasive         | 44.0  | H       | 0.68 | This study |
| South Italy    | NNP      | Not present      | 39.5  | H       | 0.68 | This study |

species’ invasiveness with regard to the northern and southern parts of Italy; and (3) evaluate the potential threats posed by the species for Europe using an impact assessment classification system. Risk analysis of non-native species, which involves risk identification (a.k.a. screening), full risk assessment, risk management and communication, is vital for the protection of European freshwater ecosystems. The outcomes of risk identification and assessments provide the evidence base with which to inform policy makers and environmental managers as to which species are likely to become invasive and exert adverse impacts on native species and ecosystems. This information assists decision makers in identifying where management resources can be most effectively deployed to eradicate or control invasive species.

**Methods**

Literature searches were undertaken via [www.google.com](http://www.google.com) and [www.google.scholar.com](http://www.google.scholar.com) using a variety of
search terms (invasive, alien, non-native, impact, aquaculture, disease, pathogens, control, management) in combination with the species’ common and scientific names. A total of 512 articles were retrieved, and of these, those containing valuable and for this study relevant information on this species’ invasiveness were retained. Also, online databases (www.fishbase.org, Froese and Pauly 2010; www.cabi.org, CABI 2016; www.gbif.org, gbif.org 2020; all accessed: 15.01.2021) and grey literature were consulted. Lastly, various ichthyologists, zoologists and experts from several European countries were consulted on their knowledge of channel catfish introductions in their respective countries. Grey literature on channel catfish was limited (e.g. Poe and Rieman 1988; Jackson and Badame 2002), so only peer-reviewed literature sources were ranked according to the number of references sourced and their reliability as: ‘low’ (one reference), ‘medium’ (multiple references or single plus personal reference) and ‘high’ (multiple recent references). The same bibliographic search strategy and sources were used to answer questions in the risk screening and impact assessments described here below. Conversions factors for fish lengths (fork and standard) into total length (TL) were taken from www.fishbase.org, whereby conversions of standard length to TL employed the mean (1.2605) of the two values given on FishBase.

Consistent with the stepwise risk analysis process (see Introduction), the species was evaluated in a two-step process: risk screening, followed by detailed impact assessment. Risk screening of the species was undertaken using the Aquatic Species Invasiveness Screening Kit (AS-ISK; Copp et al. 2016b), which: (i) complies with the with the ‘minimum standards’ (Roy et al. 2018) for the assessment of NN species under EC Regulation No. 1143/2014 on the prevention and management of the introduction and spread of invasive alien species (the IAS Regulation) (Council of the European Communities 2014); and (ii) offers assessors 32 languages with which to carry out their assessments (Copp et al. 2021). As with its parent screening tools, the Weed Risk Assessment (Pheloung et al. 1999) and the Fish Invasiveness Screening Kit (Copp et al. 2009b; Vilizzi et al. 2019), the AS-ISK comprises 49 basic (BRA) questions to evaluate a species’ biogeography/history, biology/ecology, and nuisance traits, followed by six additional Climate Change Assessment (CCA) questions, which require the assessor to estimate how the future climatic conditions are likely to affect the risks of introduction, dispersal, establishment and impacts of the assessed species (Copp et al. 2016b). For each question, the assessor must provide a response and a justification (i.e. bibliographic support) for that response, and also to rank their level of confidence in that response. As recommended by the Intergovernmental Panel on Climate Change (IPCC 2005; e.g. Copp et al. 2009b), the confidence factor (CF) rankings are: Low (2 out of 10 chance); Medium (5 out of 10 chance); High (8 out of 10 chance); Very high (9 out of 10 chance).

Using the AS-ISK’s German-language option, but providing justifications in English, the risk screenings were carried out separately for the north and northeast of Italy (henceforth, North Italy) and south and southwest of Italy (henceforth, South Italy), i.e. the two risk assessment (RA) areas because of their different climate types, according to Köppen-Geiger (Beck et al. 2018), and different freshwater ecoregions (Abell et al. 2008). The Köppen-Geiger climate types are C.fb (temperate oceanic climate) and C.sa/Csb (hot-summer Mediterranean climate) for North and South Italy, which both possess similarities with the species’ native range (Beck et al. 2018). The freshwater ecoregion 415 (Gulf of Venice Drainages) (Abell et al. 2008) corresponds to the North Italy RA area, and ecoregion 416 (Italian Peninsula and Islands) corresponds to the South Italy RA area (Supplementary Material 3). The distinction of different freshwater ecoregions within a country in risk screenings has been found to be important to ensure the most accurate evaluation of potential risk, especially within an environmental management context (Dodd et al. 2019).

In the absence of a calibrated AS-ISK threshold score for non-native freshwater fishes with which to distinguish between high and medium-to-low risk species in Italy (sensu Copp et al. 2009b), the global threshold scores acquired for temperate zone freshwater fishes in a global trial of the AS-ISK (L. Vilizzi, G.H. Copp et al. unpublished) were used (i.e. BRA = 15.9; BRA+CCA = 16.0). This global trial study includes AS-ISK assessments of channel catfish for several RA areas (Belarus, Southern Finland, Turkey, Vietnam, Anzali Wetland Complex—Iran, River Ob Basin—Russia) as well as the two assessments for north and south Italy.
Following the risk screening, the species was subjected to a more comprehensive assessment of the magnitude of channel catfish adverse impacts in the RA area using the Environmental Impact Classification of Alien Taxa (EICAT) of Blackburn et al. (2011), which was applied with regard to the species’ introduced ranges in North America and in Europe. Based on the generic impact scoring system (GISS) of Nentwig et al. (2010, 2015), the EICAT was modified to align it to the IUCN’s impact scheme (Kumschick et al. 2012) and has recently been announced (IUCN 2020) as the IUCN’s standard for the classification of non-native species impacts on the environment (www.iucn.org/theme/species/our-work/invasive-species/eicat). The system ranks the magnitude of species environmental impacts using five semi-quantitative scenarios, ranging from Minimal (species is unlikely to cause deleterious impacts on the native biota or abiotic environment) to Massive (species is likely to lead to the replacement and local extinction of native species, and to produce irreversible changes in the structure of communities and the abiotic or biotic composition of ecosystems). A species’ environmental impacts are measured using 12 classes of impact mechanism, each associated with one of a sequential series of five impact scenarios, which describe the increasing magnitude of the species’ impact. For each impact mechanism in EICAT, a confidence level is also allocated: high, medium, and low (Blackburn et al. 2011). Species are assigned to an impact category based of the largest impact value recorded, whereas the overall confidence is taken from the impact category attributed the highest confidence rank (Blackburn et al. 2011).

Results

The AS-ISK assessments of channel catfish resulted in high scores and corresponding high-risk classification for both North and South Italy (Table 1; see also Supplementary Material 4), though the scores for North Italy (BRA = 44.0) were higher than those for South Italy (BRA = 36.5). These differences were attributable to the sections A2 (Climate, distribution and introduction risk) and A3 (Invasive elsewhere). The likely impact of climate change on the screenings also differed between the north, where invasiveness risk increased (BRA + CCA = 54.0) in contrast to a decrease in the south (BRA + CCA = 28.5). Assessor confidence was relatively high for both North and South with regard to the BRA (CFs = 0.80 and 0.92, respectively), but with regard to climate-change impacts on likely invasiveness (BRA + CCA) confidence dropped less for the North (CF = 0.54) than for the South (CF = 0.25). These differences could be attributed to the sections A2 (Climate, distribution and introduction risk), A3 (Invasive elsewhere) and C9 (Climate change).

The EICAT impact assessments revealed that the species can be classified as exerting “Major impacts” in both North America and Europe, with confidence levels of “low” and “medium” for North America and Europe, respectively (Table 3). With both the AS-ISK and EICAT evaluations, the scarcity of information on the risks and impacts of introduced populations was a major contributor to the confidence rankings.

Discussion

Information available in the literature on channel catfish was mainly on the species’ environmental biology in its native, North American range, with limited information from Europe and elsewhere. Following the species’ first introduction to Europe in the 19th century for angling and aquaculture, it has been recorded in at least 22 EU-countries, being considered as established in nine (Welcomme 1988; Elvira and Almodóvar 2001; Copp et al. 2005a; Olenin et al. 2008, CABI 2016; Supplementary material 2; Fig. 1) with recent information on its potential impact reported in Italy (Haubrock et al. 2018a, b).

There have been few studies to examine the appropriate spatial scale for risk assessments: national, regional, legislatively defined (e.g. River Basin District) or cli-mo-geographic (Dodd et al. 2019). This is relevant to both non-native and translocated species, with differing climates within a region (present study) or River Basin District (Dodd et al. 2019). Differences in climate suitability have implications for the success of invasive species. Especially temperature regimes affect e.g. the success of spawning. Although the Köppen-Geiger climate in South Italy (Csb/a, at least one month’s average temperature above 22 °C) is characterised by a hotter summer than North Italy (Cfb; with all months with average temperatures below 22 °C) and therefore comparable to I.
Indeed, while the climate zone in the species’ native area North America is characterised by at least one month’s mean temperature $> 22 \, ^\circ\text{C}$, future temperature increases under recently modelled climate change scenarios (Solomon 2007; Collins et al. 2013) will likely facilitate dispersal and establishment of channel catfish *Ictalurus punctatus* in the introduced range in North America and Europe.

### Table 3

Outcomes of the environmental impact classification of Alien Taxa (EICAT; Blackburn et al. 2011) applied to the channel catfish *Ictalurus punctatus* in the introduced range in North America and Europe

| Ecological impacts                   | North America | Impact | Uncertainty | Explanation                                                                 | Europe | Impact | Uncertainty | Explanation                                                                 |
|--------------------------------------|---------------|--------|-------------|-----------------------------------------------------------------------------|--------|--------|-------------|-----------------------------------------------------------------------------|
| Competition                          | 4             | 1      |             | Competition for food, site for reproduction showed in field and laboratory experiments | 3      | 3      |             | No data of competition with native species. It can only be hypothesized from data |
| Predation                            | 4             | 1      |             | Stomach content analyses and laboratory experiments with different prey items | 4      | 2      |             | Stomach content analysis revealed high predation and opportunistic behaviour |
| Hybridisation                        | 4             | 1      |             | It can hybridize with other native Ictaluridae                              | 1      | 2      |             | No hybridization reported                                                   |
| Transmission diseases to native      | 1             | 3      |             | No data reported                                                            | 1      | 2      |             | No documented diseases transmitted to native species, but possibly to *Silurus glanis* |
| Parasitism                           | 1             | 3      |             | Nematodes are reported in the species but transmission to other species has not been documented so far | 1      | 3      |             | No impacts reported                                                       |
| Poisoning/toxicity                   | 1             | 1      |             | Spines are toxic and a defence mechanism, but no data on heavy effects on native species | 1      | 2      |             | Spines are toxic and a defence mechanism, but no data on heavy effects on native species |
| Bio–fouling                          | 1             | 1      |             | No impacts reported                                                         | 1      | 1      |             | No impacts reported                                                       |
| Herbivory                            | 3             | 2      |             | It is opportunistic and can eat algae and small aquatic plants              | 3      | 2      |             | It is opportunistic and can eat algae and small aquatic plants.            |
| Physical, chemical, structural       | 1             | 2      |             | No impacts reported                                                         | 2      | 3      |             | Only hypothesis                                                           |
| Interaction                          | 1             | 3      |             | No data reported                                                            | 2      | 3      |             | It can transmit diseases to other alien fishes, it can predate American crayfish; it can hybridize with other alien Ictaluridae |

**Final score** | 4 | **Category** | Major (MU) | 4 | **Category** | Major (MU) |

**Level of confidence** | 3 (High) | **Level of confidence** | 2 (Medium) |
catfish in North Italy, including areas of higher altitude. A similar increased distribution may also occur in South Italy, though dry summers with low precipitation will likely render some locations unsuitable. However, previous risk screenings of channel catfish for several European Countries (Greece, Bulgaria, Finland and the Iberian Peninsula) using the precursor to the AS-ISK, freshwater Fish Invasiveness Screening Kit (FISK), resulted in medium and high risk classifications—this includes Turkey for which the species was screening using both the FISK and the AS-ISK (Table 2), with the lowest risk rankings (and low scores) being for Lake Balaton (Hungary) and for southern Finland. When considering the potential impact for the European continent as indicated by EICAT, a “major impact with medium uncertainties”, corroborates the FISK and AS-ISK scores (Tables 1 and 2).

The combinations of risk screening and impact assessment for the two RA areas revealed several unstudied aspects of channel catfish environmental biology, especially pertaining to the species’ invasiveness. These gaps in knowledge become more prominent when focused on populations in Europe, as only a handful studies focused on the invasiveness and impacts of this species (see Haubrock et al. 2018a, b). Nonetheless, data on translocated populations in North America revealed that channel catfish were considered invasive where introduced, attracting the highest invasiveness ranking (based on ‘expert opinion’) for the Colorado River Basin (Hawkins and Nesler 1991), with potential impacts on endangered species, fisheries and angling. The invasiveness potential of channel catfish is enhanced by its migratory behaviour, i.e. to spawning and feeding habitats in spring and autumn, respectively (Duncan and Myers 1978; Becker 1983; Dames et al. 1989; Hanzawa and Arayama 2007). The species’ moderately broad temperature requirement (16–24 °C) to induce spawning activity (Appelget and Smith 1951) would also

Fig. 1 Map of Europe indicating countries (light grey) and the year when the channel catfish *Ictalurus punctatus* was introduced. Countries where evidence exists for the establishment of populations in the wild are highlighted with dashed lines. For Italy, the climatic zones of the temperate north and continental south are indicated.
facilitate establishment in various climatic regions globally. Similar to many invasive freshwater fishes, the channel catfish is an opportunistic, omnivorous forager, and like other ictalurid catfishes, it possesses venomous dorsal and pectoral spines, which can dissuade some predators (Supplementary Material 1). However, genetic differences between native and introduced populations can lead to variations in the expression of biological traits (Gozlan et al. 2020), which may introduce errors in the interpretation of native-range data in risk assessments—a methodological issue that demands further investigation.

Furthermore, the distribution data revealed that much information on the species’ establishment success is purely anecdotal, resulting in low confidence levels. That said, the Mediterranean distribution of channel catfish is most likely linked to angling (Banha et al. 2017), a known vector for the introduction and secondary spread on non-native fishes (e.g. Copp et al. 2005b; Kilian et al. 2012; Anderson et al. 2014). Whereas most channel catfish populations in Eastern Europe can be linked to aquaculture escapes (Savini et al. 2010), such as during natural flood or spate events (Robinette and Knight 1981; Townsend and Winterbourn 1992; Orsi and Agostinho 1999; Zanata et al. 2016). Moreover, its invasiveness is bolstered in some countries, where it is popular among anglers, and also because it can be easily reared in aquaculture. Based on information from the species’ introduced North American range, and the likelihood of accidental escapes, the proposed introduction of channel catfish to New Zealand was denied due to the high risk posed by the species to native species of conservation concern (Townsend and Winterbourn 1992).

Recorded impact mechanisms are manifold and include: (i) predation; (ii) competition; (iii) hybridisation with congeneric catfishes (Goudie et al. 1994; Dunham and Masser 2012), though none of these are present in Italy (Movchan et al. 2014); (iv) disease transmission, and (v) modification of ecosystem services and related socio-economic impacts (see Supplementary Material 4). Among the impact mechanisms the species could exert in Europe, and especially Italy, include predation on benthic species, insects, plants and macroinvertebrates, which could reduce native biodiversity. Particularly, with the high densities tending towards young age groups, behavioural interferences as well as competition for food and shelter may result in decreasing abundances of native fish (Jenkins and Burkhead 1994). However, beside native fishes, other several already endangered species may be threatened by its presence due to its opportunistic feeding activity (Endo et al. 2015; Haubrock et al. 2018a, b; 2020).

Studies of introduced pumpkinseed L. gibbosus have shown that, according to life-history traits (mortality rate, plasticity and reproductive strategy), maturity is reached earlier in Europe (Cucherousset et al. 2009). In channel catfish, which mature generally at ages 5–8 years in Canada (Scott and Crossman 1973) and 2.0–5.6 years in the southern U.S.A. (e.g. Perry and Carver 1972), though ages at maturity of < 2 years and 9 years have been reported (Bates et al. 2001; Shephard and Jackson 2005), spawning can be inhibited if suitable nesting cover is unavailable (Marzolf 1957). Furthermore, climatic differences can affect growth, spawning, and metabolism and eventually lead to precocious maturity and further unforeseeable changes in behavioural patterns altering its invasiveness (Wellborn 1988). For instance, pumpkinsseed cease feeding activity at temperatures below 8.5 °C (Keast 1968). Needing higher temperatures for feeding (> 15.5 °C, Brown 1942), channel catfish growth rate has been found to increase with increasing population density (Shephard and Jackson 2009), and growth phases are limited to temperatures > 18 °C (Starostka and Nelson 1974). Indeed, elevated water temperature can lead to faster growth, such as observed in a cooling-water reservoir in Bulgaria, where channel catfish reached TLs at age 2 Bulgaria of 59.2 cm TL (Hubenova et al. 2014); this is about twice that (26.9 mm TL) reported in Scott and Crossman (1973) for age-2 fish in the Upper Mississippi (Iowa) and far exceeds the TLs at age 2 reported in for channel catfish in various lakes (10.8–20.4 cm TL; Starostka and Nelson 1974) and rivers (11.4–19.3 cm TL; Shephard and Jackson 2006) in the U.S.A.

In both North and South Italy (see Methods), the growth season is from June to September (i.e. mean temperatures > 18 °C; World Meteorological Organization 2011). This indicates that channel catfish should be able to undertake growth throughout Italy, but reproduction may be expected to be limited to waters in which mean temperatures reach 23.9–29.5 °C (Scott and Crossman 1973).

Similar to other (non-native) ictalurid catfish population in Europe, the channel catfish has a limited
number of natural predators, which include pikeperches Sander spp. (Scott and Crossman 1973; Hanchin et al. 2002), cormorants Phalacrocorax spp., herons Ardea spp., pelicans Pelecanus spp. and otters (Lutra lutra). Younger (smaller) channel catfish are particularly susceptible to avian predators (Glahn and Coble 1971) and field studies of northern pike and walleye Sander vitreum diet found channel catfish to be taken by walleye but not northern pike (Tyus and Beard 1990). This information from the native range suggests that the main potential predators of channel catfish in Europe are otters, herons and cormorants, pikeperches and European catfish Silurus glanis (Copp et al. 2009a).

Nevertheless, potential predation by native or non-native species has yet not been assessed, though diet analysis of European catfish has found this species to predate small (< 80 cm TL), immature individuals of channel catfish (Haubrock et al. 2019) and black bullhead of 105–200 mm TL (Pouyet 1987; see also Copp et al. 2009a). Angling of channel catfish is lower in Italy than in the native region (Leonard et al. 2010), suggesting that little if any human pressure is exerted on this species. However, human alteration of the environment has long been reported to facilitate species invasions (Moyle 1986; Havel et al. 2005; Olden and Poff 2005; Cucherousset et al. 2009, 2012).

A very well-known example of another North American catfish invader in Europe, including in Italy, is the black bullhead, which first appeared in the United Kingdom in 1885 (Wheeler 1978), reaching Italy in 1904 (Tortonese 1970). The key drivers for non-native fish introductions have often been cultural, because the stocking of exotic species is often perceived as favourable and beneficial for local municipalities and their fisheries (Selge et al. 2011; Kilian et al. 2012). These are seen as enhancements of fish stocks for recreational angling and/or aquaculture. As a result, the black bullhead has already spread throughout Italy, reaching high densities in the River Tiber and the Corbara Reservoir (Pedicillo et al. 2008). Currently, black bullhead is considered established in both North and South Italy, though populations are more common and more abundant in the north of Italy, contrasting the South, where black bullhead populations are limited to reservoirs and other drought-resistant water bodies (Nocita and Zerunian 2007; Marrone and Naselli-Flores 2015). Although black bullhead generally prefers a temperate/mesothermal climate (REF), which contrasts the tropical to mega-thermal climate preferred by channel catfish (Cheetham et al. 1976), climatic conditions may not be limiting factors for either species in Italy, but climatic conditions may be the underlying differences in the establishment success and expansion across Europe, such as reported for European catfish.

Another North American catfish present in Europe is the brown bullhead, which was found in a recent full risk assessment (for the EU as the RA area) to pose a medium risk at a moderate confidence level (Aislabie et al. 2019). Unlike the black bullhead, the brown bullhead has received limited study in Europe until recently (Kozic and Vilizzi 2021).

In conclusion, most of the information on the impacts of channel catfish is available for the species’ introduced range in North America, whereas for Europe, data are scarce (Supplementary Material 1 and 4). Risk screenings and assessments of the channel catfish have resulted in risk rankings of medium to high risk (Table 2; Supplementary Material 4), and assessment of environmental and economic impacts using the GISS resulted in a medium risk score (Van der Veer and Nentwig 2015). This contrasts the moderately-low risk ranking attributed to channel catfish using the Organism module of the ENSARS, the European Non-native Species in Aquaculture Risk Assessment Scheme (Copp et al. 2016c), though the ENSARS scoring system has yet to be calibrated. These outcomes, combined with the assessments presented here, highlight the urgent need of more research on the species in Europe in order to better investigate its potential invasiveness and promote adequate management plans. Nevertheless, the channel catfish can be considered as potentially invasive fish species in many parts of Europe, though this comes with a caveat for low certainty due to the scarcity of data on European populations and any realised impacts. Information regarding impacts were inferred from literature related to North American invaded areas and, for some impact mechanisms (e.g. predation), from direct observations (Haubrock et al. 2018b, 2019). The high physiological tolerance and fecundity accompanied by parental care make this
species a good candidate for future successful invasions throughout European water bodies (Marchetti et al. 2004). By natural or human mediated spread, urban areas with rivers already disturbed and invaded by other species are prone to introductions and subsequently facilitate its population growth. With only limited scientific attention being paid to the spread of this potentially invasive species, there are risks of at least moderate-to-major impacts to European freshwater ecosystems. In view of the presence of channel catfish in fresh waters of more than 20 European countries, the potential impacts of this species require urgent study so as to inform risk analysis of this species. Given that the AS-ISK is compliant with the ‘minimum standards’ for assessing risks under the EU IAS Regulation on (European Union 2014), the high-risk ranking found in the present risk screening suggests this species be subject to regulation under that piece of legislation. The core of the IAS Regulation is the list of Invasive Alien Species of Union concern, which is being updated as appropriate (first update on 2.08. 2017, second update 15.08.2019,) and currently third update is in process of finalisation that took into consideration two ictalurid species, *Ameirus melas* and *Ameirus nebulosus*. Species on the list are subject to the restrictions on keeping, importing, selling, breeding and growing and Member States are required to manage these species.

Control or eradication methods do currently exist for ictalurid catfishes (e.g. Copp et al. 2016a; Aldridge et al. 2019), though unlike the UK, use of rotenone elsewhere in Europe may not be possible.

**Acknowledgements** We thank K. Magellan for helpful comments on an earlier version of the manuscript and T. Júza, K. Jakubčinová, R. Striekworld, E. Uzunova, J. Cucherousset, G. Hotos, T. Friendrich and F. Ribeiro for information on the introduction history of channel catfish. This project has received funding from the European Union’s Horizon 2020 research and innovation programme under the project Marie Skłodowska-Curie Aquainvad-ED (H2020-MSCA-ITN-2014-ETN-642197). The participation of G.H. Copp was supported by Cefas’ Science Excellence fund.

**Funding** Open Access funding enabled and organized by Projekt DEAL.

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