Analysis of the structural characteristics and spatial organization of macrobenthic fauna in Oualidia lagoon, Morocco

Fatima El Asri1,5 · Ahmed Errhif1 · Mohamed-Naoufal Tamsouri2 · Daniel Martin3 · Mohamed Maanan4 · Hakima Zidane5

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Abstract
Morocco is a very rich country in terms of wetlands, including bays, estuaries, lagoons, lakes, and rivers, among others. However, many of them need to be well managed and better exploited, thus requiring baseline descriptive studies to assess faunal patterns and trends. Oualidia lagoon is not an exception. In fact, the studies conducted to date in the lagoon mainly focused on its oceanographic features and aquaculture capacities. The present paper presents the first study on the whole macrofaunal assemblages of Oualidia lagoon, having as main objectives analyze the composition and structure of the assemblages, as well as their relationships with the main environmental variables explaining their spatial distribution variability in the enclosed ecosystem of the Oualidia coastal lagoon. The benthic macroinvertebrates of Oualidia lagoon were studied in summer 2013 based on 43 stations spread over the whole lagoon by using a Van Veen grab. Two replicate samples were collected for macrobenthos and one additional replicate for sediment analyses at each station. Also, single water salinity and temperature measurements were recorded at each station. Fifty-one species from seven different taxa were recorded, among which mollusks (37%), polychaetes (25%) and crustaceans (25%) dominated in terms of species number. Peringia ulvae, Abra alba, Sphaeroma serratum, Cerastoderma edule and Corophium sp. accounted for 88% of total abundance. The benthos was structured along a downstream–upstream gradient, showing three assemblages, named according to the IndVal index: (1) Diopatra cf. marocensis, (2) Peringia ulvae, and (3) Cerastoderma edule, while the main factors driving their distribution were granulometry and salinity. The Biotic and Environmental analysis showed that the combination of granulometry and salinity was the major factors controlling the spatial distribution of the macrozoobenthos in the Oualidia Lagoon.

Keywords Benthic macrofauna · Biodiversity · Distribution · Community structure · Environmental factors · Oualidia lagoon · Morocco

Introduction
Coastal lagoons are highly productive ecosystems acting as feeding, nursery, reproduction, and protection spots for invertebrates, fishes, and birds, as well as rich fishery grounds (Newton et al. 2018). However, they are fragile ecosystems, highly susceptible to both natural and anthropogenic disturbances (Blanchet et al. 2008; Rossi et al. 2006; Affian et al. 2009).

Many interactions occurring in coastal lagoons are mediated through macrobenthic organisms which: (1) contribute to mineralize, promote, and mix the oxygen flux into the sediment, (2) recycle the organic matter and nutrients, (3) decompose detrital matter, (4) stabilize sediments, and (5) contribute to the trophic web by linking producers to higher consumers such as fishes, birds and epibenthic crustaceans (Pearson and...
Rosenberg 1978; Nybakken 1993; Heilskov and Holmer 2001; Newton et al. 2018). Such a relevant role, together with their sensitivity to human disturbance, leads these organisms to be often employed as indicators of coastal ecosystems health, and coastal lagoons are not an exception (Dauer 1993; Borja et al. 2000; Salas et al. 2004). The relatively limited capacity of dispersion and sedentary habits of adult benthic organisms does not allow them to easily avoid unfavorable environmental conditions. Consequently, they are considered as sensitive indicators of both naturally and anthropogenically driven changes in coastal ecosystems (Salas et al. 2004; Borja et al. 2000). Moreover, their relatively long lifespan allows them to integrate water and sediment quality conditions through time, as well as to show responses to temporal and chronic disturbances. Particularly, both stress tolerant and sensitive benthic species may show changes in diversity, biomass and abundance, while their assemblages may show changes in trophic and functional structure (Pearson and Rosenberg 1978; Reiss and Kröncke 2005).

The above-mentioned relationships between benthic organisms and environmental conditions in coastal lagoons have been seldom analyzed along Moroccan coasts (El Asri et al. 2018, 2021). Nevertheless, the country has a rich network of lagoons, particularly in the Atlantic littoral where Oualidia lagoon is located. Oualidia is the most important wintering area for migratory birds in Morocco (El Hamoumi et al. 2003) and a Wetland of International Importance (under the RAMSAR Convention). It also provides a valuable environment for a rich variety of plants, fish, and other wildlife (El Hamoumi et al. 2003) and is currently considered a Natural Park. Moreover, the lagoon watershed supports intensive agricultural, cattle rearing, shell-fishing and tourism activities being thus an on-going economical pole (Maanan et al. 2017). Despite its intrinsic interest, most previous studies in Oualidia focused on the assemblage structure and distribution

\section*{Materials and methods}

\subsection*{Study area}

Oualidia lagoon, located in the Moroccan Atlantic Ocean coast (32°44'42" N, 9°02'50" W, Fig. 1), measures 7 km long and about 1 km wide. Characteristically, it shows numerous side channels connected to a meandering main channel with 2 m of average depth and 5 m of maximum depth during flood tides (Bidet and Carruesco 1982). Flood tides cover more than 75% (2.25 km²) of the lagoon surface, bringing salt water to its upstream reaches and into a saline marsh beyond its second dam (Maanan et al. 2014; Hilmi et al. 2017). Salinity ranges from 30 to 36% in the inner areas with very low mixing and very weak stratification of the water column (Makaoui et al. 2018). The rainfall reached c.a. 250 mm per year, and the prevailing winds are W–NW and E, almost parallel to the major lagoon axis (Zourarah et al. 2007).

\subsection*{Sampling and data analysis}

Forty-three stations were sampled in July 2013 (Fig. 1) using a Van Veen grab (0.125 m² surface area). Two replicate samples were collected for macrobenthos and one additional replicate for sediment analyses at each station. Also, single water salinity and temperature measurements were recorded at each station with the help of a thermo-salinometer. The macrobenthos replicates were sieved in situ through a 1 mm pore-sized mesh, and the retained materials were transferred to plastic containers and fixed in a 10% formalin solution. The sediment replicates were transferred to plastic contained and kept cold until analyzed.

In the laboratory, the macrofauna was sorted under a binocular microscope, identified at the lowest taxonomic level possible, counted and classified into trophic groups (e.g., Fauchald and Jumars 1979; Hily and Bouteille 1999; Pranovi et al. 2000; Afli et al. 2008; Khedhri et al. 2015) as follows: (1) herbivores (H), algae-feeding organisms (e.g., some echinoids), (2) scavengers, feeding on carrion deposited on the bottom (essentially gastropods and decapods), (3) detritus feeders, feeding on particulate organic matter, essentially vegetable detritus (mainly amphipods and tanaids), (4) carnivores, predators like errant polychaetes or sea-anenomes, (5) suspension feeders, feeding on food suspended in the water column (e.g., most bivalves and some polychaetes), (6) surface deposit feeders, feeding on organic particles settling on the sediment (most
sedentary polychaetes), and (7) subsurface deposit feeders, burrowers that ingest the sediment from which they extract their food.

Sediment replicates were used to measure: (1) chlorophyll ‘a’ contents (mg/m², Lorenzen method, Holm-Hansen et al. 1965), (2) percentages of the total organic matter (TOM, % of weight loss on ignition 4 h at 450 °C), and (3) grain size (µm, laser granulometer Malvern, Mastersizer).

Macrobenthic spatial distribution and biodiversity were described by the species richness (S), abundance (A), Shannon diversity (H′, as \( \log_{2} \)) (Shannon 1948) and evenness (J′) (Pielou 1966), as estimated with the Paleontological Statistics program (PAST v2.14) (Hammer et al. 2001). The structure of the assemblages was analyzed by a Hierarchical Ascending Classification (HAC) on a density [transformed to \( \log_{10} (x + 1) \) to limit the influence of the most dominant taxa] per station matrix using the Bray–Curtis distance and the Ward’s method. Each identified assemblage was classified according to the IndVal index (Dufrêne and Legendre 1997). Visually observed patterns and correlations of species density with abiotic parameters were quantitatively explored with the BIOENV (Biotic and Environmental linking) routine, based on Euclidean distances (for the abiotic dataset) and Spearman’s rank correlations. All multivariate analyses were executed on an R environment (www.r-project.org). The results were plotted on maps using ARCGIS 10.0 (Figs. 2, 3).

Results

Macrofaunal descriptors

A total of 51 species of macrobenthic invertebrates were identified (Table 1), of which mollusks were the most diverse group (19 species, 37%), followed by polychaetes and crustaceans (13, 25.5% each) and echinoderms (three, 5.9%). Cnidarians, insects and tunicates (one species each) constituted a 6.0%. Mollusks clearly dominate in abundance (77% of all specimens recorded), followed by far by crustaceans (15%), polychaetes (5%), and echinoderms, cnidarians, insects and tunicates (3% altogether).

Density ranged from 0 to 6336 individuals per m² (1274 ± 1469 ind. m⁻², mean ± SD). The stations showing the highest densities were roughly situated in the center of the lagoon (Fig. 4a). The number of species per station ranged from 0 (station 7) to 18 (station 40), with 7.4 ± 3.7 species (mean ± SD). The stations showing the highest
The “best combination” of abiotic parameters explaining the observed density patterns was obtained by combining granulometry and salinity (BIOENV, \( R = 0.49 \)) (Table 2).

**Assemblage structure**

Three groups of benthic organisms were identified in the lagoon (HAC, Figs. 6, 7). Group 1 included six stations from the outer part, close to the inlet. It was characterized by sediments with median-coarser sand with low organic matter content, which were subject to strong tidal currents. The total number of species was 43, mainly distributed among mollusks (37%), crustaceans (28%) and polychaetes (28%) (Fig. 8a). The average density was low \( (529 \pm 475 \text{ ind.m}^{-2}) \) comparing to other groups, and the dominant taxa were mollusks (58%) and crustaceans (25%) (Fig. 8b), while the average evenness was \( 0.7 \pm 0.2 \) (Table 3). The assemblage was dominated in density by both surface deposit and suspension feeders (76.1%) (Fig. 9a), while carnivores (22%), detritus feeders (20%), surface deposit feeders (18%) and suspension feeders (18%) dominate in species richness (Fig. 9b). The most dominant species were the gastropod *Peringia ulvae* (31.7%), the crustacean *Corophium* sp. (24%), followed by far by the bivalve *Abra alba* (18%).

Group 2 included 21 stations mainly located inwards, along the middle area of the lagoon, and characterized by having mixed silty-sand and sandy-silt sediments. The average density was \( 1712 \pm 1644 \text{ ind.m}^{-2} \), and mollusks were largely the most dominant, with more than 90% of the total (Fig. 8b). There were 25 species, most of which belonging to crustaceans (11.4%) and mollusks (9.36%) (Fig. 8a) and the evenness was low \( (0.5 \pm 0.2) \) (Table 3). The assemblage was dominated in density by surface deposit feeders (ca. 80% of the total density), while other trophic groups were present with very low densities (Fig. 9b). Surface deposit feeders also dominate in number of species (24%), but were...
| Taxonomic groups | Family | Species |
|------------------|--------|---------|
| Mollusca         | Cardiidae | *Cerastoderma edule* |
|                  | Donacidae | *Donax trunculus* Linnaeus, 1758 |
|                  | Semelidae | *Abra alba* (W. Wood, 1802) |
|                  | Lucinidae | *Loripes orbiculatus* Poli, 1791 |
|                  |            | *Lucinoma borealis* (Linnaeus, 1767) |
|                  | Mactridae  | *Spisula solida* (Linnaeus, 1758) |
|                  | Veneridae  | *Ruditapes decussatus* (Linnaeus, 1758) |
|                  | Mytilidae  | *Mytilus galloprovincialis* Lamarck, 1819 |
|                  | Haminocidae | *Haminoea cf. japonica* Pilsbry, 1895 |
|                  | Aplysidae  | *Aplysia punctata* (Cuvier, 1803) |
|                  | Onchidiidae | *Onchidella celtica* (Cuvier, 1817) |
|                  | Naticidae  | *Cochlis vittata* (Gmelin, 1791) |
|                  | Hydrobiidae | *Peringia ulvae* (Pennant, 1777) |
|                  | Nassariidae | *Tritia pfeifferi* (Philippi, 1844) |
|                  | Trochidae  | *Tritia reticulata* (Linnaeus, 1758) |
|                  | Naticidae  | *Cymbula safiana* (Lamarck, 1819) |
| Polychaeta       | Nereididae | *Hediste diversicolor* (O.F. Müller, 1776) |
|                  | Polyonidae | *Harmonia sp.* |
|                  | Phyllodocidae | *Phyllophora sp.* |
|                  | Glyceridae | *Glycera cf. tridactyla* Schmarda, 1861 |
|                  |            | *Glycera alba* (O.F. Müller, 1776) |
|                  | Nephtyidae | *Nephtys kersivalensis* McIntosh, 1908 |
|                  | Lumbrineridae | *Lumbrineris coccinea* (Renier, 1804) |
|                  | Onuphidae  | *Diopatra cf. marocensis* Paxton, Fadlaoui & Lechapt, 1995 |
|                  | Ampharetidae | *Ampharetia romijni* Horst, 1919 |
|                  | Pectinariidae | *Lagis cf. koreni* Malmgren, 1866 |
|                  | Capitelliidae | *Capitella sp.* |
|                  | Sabellidae  | *Panousea africana* Rullier & Amoureuse, 1969 |
| Crustacea        | Melitidae  | *Melita palmata* (Montagu, 1804) |
|                  | Corophiidae | *Corophium sp.* |
|                  | Caprellidae | *Caprella liparotensis* Hailer, 1879 |
|                  | Amphipodidae | *Ampiochus sp.* |
|                  | Anthuridae  | *Cyathura carinata* (Krøyer, 1847) |
|                  | Sphaeromatidae | *Sphaeroma serratum* Fabricius, 1787 |
|                  | Idoteidae  | *Idotea balitica* (Pallas, 1772) |
|                  | Tanaididae | *Tanaiss dulongii* (Audouin, 1826) |
|                  | Apseudidae | *Apseudes sp.* |
| Cumaecae         | Pagurididae | *Pagurus bernhardus* (Linnaeus, 1758) |
|                  | Portunididae | *Carcinus maenas* (Linnaeus, 1758) |
|                  | Balanididae | *Balanus sp.* |
| Insecta          | Chironomidae | *Chironomus sp.* |
| Echinodermata    | Holothuridae | *Holothuria polii* Delle Chiaje, 1824 |
|                  | Ophiuridae  | *Ophiura sp.* |
|                  | Parechinidae | *Paracentrotus lividus* (Lamarck, 1816) |
| Cnidaria         | Hormathiidae | *Calliactis parasitica* (Couch, 1842) |
| Chordata         | Ascidiacea  | *Ascidia sp.* |
closely followed by detritus feeders (22%) and scavengers (18%) (Fig. 9a). The most dominant species were the gastropod *P. ulvae* (66%), followed by far by the bivalve *A. alba* (22%).

Group 3 included 16 stations mainly located in the most inner area of the lagoon and characterized by variably silty sediment with high organic matter content (in shallower bottoms) and muddy sediments (in deeper bottoms).
The total number of species was 25, which were distributed mainly among crustaceans, mollusks and polychaetes (Fig. 8a). The mean density was $1,643 \pm 1,641$ ind.$m^{-2}$, with the mollusks being the most dominant (62%), followed by crustaceans (31%). All remaining groups just represented a $\leq 2\%$ (Fig. 8b). The evenness was $0.7 \pm 0.1$ (Table 3). The assemblage was dominated in density by surface deposit feeders (44.7%), followed by detritus feeders (26.3%) and suspension feeders (25.8%) and the other groups ($\leq 3\%$) (Fig. 9b). Conversely, the number of species detritus feeders, suspension feeders and surface deposit feeders is 24% each (Fig. 9a). The most dominant species were *A. alba* (24.2%), *Sphaeroma serratum* (21.3%), *P. ulvae* (21%), *Cerastoderma edule* (16.9%) and *Corophium* sp. (4%).

**Discussion**

The present study represents a reliable baseline for future ecological research by providing a detail description of the observed spatial variability in summer conditions, as well as the relationships between the environmental factors, the biotic descriptors and the assemblage structure. The water temperature increased from outer to inner parts of the lagoon due to the influence of the cold ocean waters entering through the inlet (Fig. 2a). The highest salinities are observed mainly at the stations located downstream of the lagoon, directly subject to marine influences. The low values recorded upstream of the lagoon are due to the resurgences of fresh water which are responsible for the slight desalination of the lagoon water upstream. The low value observed at station 43 is explained by the dilution by fresh water from nearby water sources. Both granulometry (Fig. 3c) and organic matter (Fig. 3a) patterns mirrored that of temperature, revealing the relevant influence of the hydrodynamic functioning of the lagoon, with the inner area tending to be relatively calm, thus to having more fine sediments and organic matter (Hilmi et al. 2005; Maanan et al. 2014). Conversely, salinity decreased from outer to inner parts (Fig. 2b), where there were the main freshwater arrivals (Hilmi et al. 2005; Damsiri et al. 2014; Hassou et al. 2014). Moreover, also, a minimum of salinity downstream was detected (but in a single station), which was probably caused by the numerous underwater freshwater outbreaks located in this area of the lagoon (Rharbi et al. 2001; Hilmi et al. 2005).

This environmental framework characterizes Oualidia as a typical lagoon environment, where the benthic macrofauna is dominated by gastropods, bivalves, amphipods, isopods and polychaetes (Mistri 2002; Bazaïri et al. 2003). However, the presence of cnidarians, tunicates and insects was also recorded. The former two occurred likely in connection with the seaweed coverage, while the later was only present in the most enclosed parts of the lagoon, where the salinity was low due to the numerous underwater freshwater outbreaks.

The biodiversity in Oualidia largely doubled the previous record in the area by Chbicheb (1996), 51 versus 24, and was also higher than those reported in nearby lagoons (Mergaoui et al. 2003; Marchini et al. 2004; Chaouti and Bayed 2005). This relatively higher number of species may be related to the combined influence of the open sea at the entrance, the underground freshwater seepage in the central area and the main freshwater inputs in the most enclosed area (Carruesco 1989; Hilmi et al. 2005; El Asri et al. 2015, 2018). This particular structure seemed to favor a mixed presence of marine,
Our results based on the whole macrofauna in summer agree with the previously reported structuring based on mollusk (El Asri et al. 2015) and polychaete assemblages in summer and winter in Oualidia (El Asri et al. 2018) when were studied independently and, to some extent, with those of the whole macrofauna in winter (El Asri et al. 2021). In all cases, there were three assemblages clearly organized along the outer–inner axis of the lagoon. As reported for other lagoons, this type of assemblage structuring resulted from combining the particular habitat preferences of all macrobenthic species (Bazaïri et al. 2003; Lakhdar Idrissi et al. 2004; Chaouti and Bayed 2008; Rodrigues et al. 2011; Lefrere et al. 2015). However, lagoon environments generally show marked seasonal variability (e.g., Bertrán et al. 2016), and there are no reasons allowing us to suspect that Oualidia could be an exception. Therefore, to assess the possible seasonal differences in the macrofaunal assemblages of Oualidia, there is an on-going research attempting to compare the summer situation analyzed in the present work with the results of a previous study carried out in winter conditions (El Asri et al. 2021).

Table 3 List of the main species of each macrofauna community classified by descending order using the IndVal index values

| Cluster | Species                | IndVal |
|---------|------------------------|--------|
| 1       | Diopatra cf. marocensis| 196.8  |
|         | Lumbrineris coccinea   | 92.3   |
|         | Loripes orbiculatus    | 82.8   |
|         | Tritia reticulata      | 70.8   |
|         | Ruditapes decussatus   | 52.9   |
| 2       | Peringia ulvae         | 190.5  |
|         | Tritia pfeifferi       | 129.6  |
|         | Hediste diversicolor   | 127.9  |
|         | Abra alba              | 127.7  |
|         | Calliactis parasitica  | 64.4   |
| 3       | Cerastoderma edule     | 257.4  |
|         | Melita palmata         | 232.8  |
|         | Sphaeroma serratum     | 163.4  |
|         | Cyathura carinata      | 139.5  |
|         | Abra alba              | 132.9  |

Fig. 8 Faunal composition structure of the macrofauna community expressed in relative number of species (a) and individuals (b).

Fig. 9 Trophic structure of the macrofauna community expressed in relative number of species (a) and individuals (b). Surface deposit feeders (SDF), Subsurface deposit feeders (SSDF), Detritus feeders (Dt), Suspension feeders (S), Carnivores (C), Scavengers (N), Herbivores (H)

brackish and freshwater species in a lagoon environment (Zabi and Lelouef 1993).
In Oualidia, the benthic assemblages conforming the three groups identified in the HAC were named according to the IndVal index (Table 3). The *Diopatra cf. marocensis* assemblage (Group 1) was mainly distributed in the outer part of the lagoon and mollusks, crustaceans and polychaetes were almost equally dominant. The species richness was the highest found in the lagoon, due to the presence of both lagoon and marine species, the later entering the lagoon environment thanks to the proximity to the inlet (Cherkaoui 2006; Koaudio et al. 2008). Such a proximity also influenced the assemblage density, with the low numbers being likely caused by the high hydrodynamics favoring sedimentary instability (El Asri et al. 2015). Nevertheless, the assemblage revealed to be well-balanced, as indicated by a relatively high evenness.

The *Peringia ulvae* assemblage (Group 2) was the largest in number of stations and occurred mainly in the middle area of the lagoon. The hydrodynamic regime was moderated and the environment showed a relatively high nutrient enrichment (Rharbi et al. 2001; Hilmi et al. 2005). These conditions favored some of the established species, which led this assemblage to show the maximum average density and facilitated species enrichment but also in the maintenance of a relatively stable environment allowing the establishment and organization of a more structured community. The central zone, on the other hand, has the lowest values of these structure indices. It is a transition zone toward the interior of the lagoon where the environment is relatively unstable. The macrobenthic community at this level is unstructured and strongly dominated by only two species (*Peringia ulvae* and *Abra alba*) which are described as surface deposits. The upstream zone has the characteristics of a confined zone where the number of species is reduced but with appreciable diversity and above all fairness. The reduction in the specific richness in this area is the consequence of a selection of species by this particular environment resulting in a stable and characteristic community of the area. Similar results have been obtained in other lagoons such as the Ria d’Aveiro lagoon in Portugal (Garnerot et al. 2004), the Sacca di Goro lagoon in northeastern Italy (Mistri et al. 2001), or even in Dypso fjord in Denmark (Garnerot et al. 2004). At the level of the Bizerte lagoon in Tunisia, Belkhodja-Mahjoud et al. (2007) distinguished two different zones, one well structured with high diversity and regularity and another less structured where these indices are low reflecting a pronounced
imbalance benthic populations. The work of Chaouti and Bayed (2005) showed weak structure indices over the whole of the Smir lagoon in Morocco, testifying to a weak structure of the benthic fauna throughout this lagoon.

The weak structuring of benthic communities observed in general in paralic environments is the direct cause of the importance of variations in abiotic factors such as salinity (Llanso et al. 2002; Teske and Wooldridge 2003; Giménez et al. 2005; Pasquaud 2006), temperature and hydrodynamics (Stora et al. 1995; Blanchet et al. 2005; Tlig-Zouari et al. 2008). Added to this is the sediment factor which plays an important role in the structuring of endogenous benthic communities (Gray 1981; Mannino and Montagna 1997; Teske and Wooldrijde 2001; Ysebaert and Herman 2002). Other biotics such as the trophic environment can also intervene in the structural organization of benthic macrofauna (Solidoro et al. 2004).

The results suggest that the spatial patterns of the macrobenthic assemblages of Oualidia were mainly influenced by the salinity gradient and the sediment composition, with hydrodynamics also playing a relevant role. This agrees with previous results found in different lagoons and enclosed bays, such as the Epe Lagoon (Uwadiae 2013), the Guanabara Bay (Mendes et al. 2007), or the Paranagua Bay (Boehs et al. 2004). However, the environmental factors controlling species distribution in different paralic environments may vary. In the Smir lagoon (Chaouti and Bayed 2008), the vegetation and grain size were the major factors influencing the structure of species. In the Sacca di Goro (Mistri et al. 2001), the main control was attributed to dissolved oxygen, temperature and salinity. In the Monolimnii lagoon (Kevrekidis 2004), the most influencing factors were temperature and depth. In the Ria de Aveiro (Rodrigues et al. 2011), the spatial distribution of the fauna was found to be mainly related to the hydrodynamic regime and the salinity gradient, while the sediment grain size was much less important. Conversely, in the Arcachon Bay (Do et al. 2011), the structure of benthic communities was largely influenced by the combination of sediment grain size distribution and seagrass presence. Concerning the vegetation found at the level of the lagoon, it is the halophytic phanerogams characteristic of brackish water "the eelgrass", which predominate the intertidal zone and at the level of the schorres of the lagoon of Oualidia, as well as the green algae "the seaweeds". ulvae."

The phytoplankton is abundant and represented by five large groups (Bennouna 1999); indeed, surely, the vegetation has an effect on the distribution of the benthic macrofauna; however, it has not been studied and its impact pending to be determined.

All these paralic environments differ considerably from one another in morphology and water depth, which certainly gave rise to differences in hydrodynamic regime, temperature, salinity, granulometry, vegetal coverage, etc. This may certainly explain the reported changes in the environmental factors controlling species distribution, but also the differences in the composition and abundance of the established macrobenthos (Hernandez-Guevara et al. 2008). Moreover, the existing interactions between the macrobenthic fauna and other members of the ecosystem, including different components of the benthic compartment, but also all organisms inhabiting the lagoon (e.g., fish, birds, etc.) are certainly affecting the reported differences, particularly through predation, competition, etc. (Gambi et al. 1998; Nicolaïdou et al. 2006; Schückel et al. 2015; Khedhri et al. 2017). Finally, one cannot discard the anthropogenic influences, which tend to be a major force in driving the functioning of most paralic environments all around the world (e.g., Vezzone et al. 2020; Fang et al. 2020) and, certainly, Oualidia is not an exception (Jayed et al. 2015; Lakhlalki el al. 2017).

**Conclusion**

This study on the analysis of the structural characteristics and spatial organization of macrobenthic fauna in Oualidia lagoon has contributed to 54,407 macrozoobenthic organisms from 51 species in Oualidia, with the gastropod Peringia ulvae, the bivalves Abra alba and Cerastoderma edule, the isopod Sphaeroma serratum and the amphipod Corophium sp. being the most abundant species. Overall, the macrobenthos was structured in three assemblages distributed along an inner–outer axis, with the observed patterns in assemblage composition and structure being mainly driven by granulometry and salinity, with the hydrodynamics also playing a relevant role. This study must be considered as a baseline for future monitoring programs on the coastal management in the Oualidia lagoon. Therefore, further investigations are needed to see the whole picture of biodiversity status of macrofauna in the Oualidia lagoon, and to monitor the ecosystem health, it would be very interesting to study the seasonal variability of the benthic community in the lagoon and its linkage to the fluctuating physical–chemical parameters. To maintain high biodiversity, it is necessary to develop a specific management plan that will take into account the importance of the shallow water habitats in maintaining the ecological functions of the lagoon including that of nursery for juvenile fishes.

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Data availability  The data that support the findings of this study are available from the corresponding author upon reasonable request.

Declarations

Conflict of interest  The authors declare that they have no conflict of interest.

Ethical approval  All applicable international, national, and/or institutional guidelines for animal testing, animal care and use of animals were followed by the authors.

Sampling and field studies  All necessary permits for sampling and observational field studies have been obtained by the authors from the competent authorities and are mentioned in the acknowledgements, if applicable.

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