Spatiotemporal Variation and Influencing Factors of TSP and Anions in Coastal Atmosphere of Zhanjiang City, China

Ji-Biao Zhang 1,2*, Yu-Mei Rong 1, Qi-Feng Yin 1, Peng Zhang 1,*, Li-Rong Zhao 1 and Chun-Liang Chen 3

1 College of Chemistry and Environmental Science, Guangdong Ocean University, Zhanjiang 524088, China; zhangjb@gdou.edu.cn (J.-B.Z.); rongymeimuit.edu.cn (Y.-M.R.); yinfq@mail.sustech.edu.cn (Q.-F.Y.); zhaolr@gdou.edu.cn (L.-R.Z.)
2 Southern Laboratory of Ocean Science and Engineering (Guangdong Zhanjiang), Zhanjiang 524088, China
3 Analytical and Testing Centre, Guangdong Ocean University, Zhanjiang 524088, China; chencl89@gdou.edu.cn
* Correspondence: zhangpeng@gdou.edu.cn; Tel.: +86-0759-2383300

Abstract: Water-soluble anions and suspended fine particles have negative impacts on ecosystems and human health, which is a current research hotspot. In this study, coastal suburb, coastal urban area, coastal tourist area, and coastal industrial area were explored to study the spatiotemporal variation and influencing factors of water-soluble anions and total suspended particles (TSP) in Zhanjiang atmosphere. In addition, on-site monitoring, laboratory testing, and analysis were used to identify the difference of each pollutant component at the sampling stations. The results showed that the average concentrations of Cl\(^{−}\), NO\(_3^{−}\), SO\(_4^{2−}\), PO\(_4^{3−}\), and TSP were 29.8 µg/m\(^3\), 19.6 µg/m\(^3\), 45.6 µg/m\(^3\), 13.5 µg/m\(^3\), and 0.28 mg/m\(^3\), respectively. The concentration of Cl\(^{−}\), NO\(_3^{−}\), PO\(_4^{3−}\), and atmospheric TSP were the highest in coastal urban area, while the concentration of SO\(_4^{2−}\) was the highest in coastal industrial area. Moreover, there were significantly seasonal differences in the concentration of various pollutants (p < 0.05). Cl\(^{−}\) and SO\(_4^{2−}\) were high in summer, and NO\(_3^{−}\) and TSP were high in winter. Cl\(^{−}\), SO\(_4^{2−}\), PO\(_4^{3−}\), and TSP had significant correlations with meteorological elements (temperature, relative humidity, atmospheric pressure, and wind speed). Besides, the results showed the areas with the most serious air pollution were coastal urban area and coastal industrial area. Moreover, the exhaust emissions from vehicles, urban enterprise emissions, and seawater evaporation were responsible for the serious air pollution in coastal urban area. It provided baseline information for the coastal atmospheric environment quality in Zhanjiang coastal city, which was critical to the mitigation strategies for the emission sources of air pollutants in the future.

Keywords: coastal atmosphere; anions; TSP; influencing factor; Zhanjiang

1. Introduction

After decades of rapid economic development in China, people’s living standards have been generally improved, and demand for improving air quality is getting higher and higher. The main substances affecting air quality are coarse particulate matter and aerosol particles contained in the atmosphere. Atmospheric aerosols are gas dispersion systems in which fine particles, such as solids and liquids in the atmosphere, are dispersed in a gas and contain various substances [1]. Soluble anions (chloride ion, nitrate ion, phosphate ion, sulfate ion) and suspended fine particles are important environmental factors affecting atmospheric environmental quality. Atmospheric anions are converted mainly from their gaseous precursors, such as HCl, NO\(_X\), HNO\(_3\), HNO\(_2\), SO\(_2\), and liquid nitric acid mist and sulfuric acid mist. Previous studies have shown that the accumulation of atmospheric sulfate, nitrate, and PM\(_{10}\) levels can cause a variety of diseases, seriously affecting human health [2–8]. In recent years, researches on air pollutants focused on the temporal and spatial distribution characteristics, sources, and impact mechanisms of PM\(_{2.5}\).
SO₂, NO₂, NH₃, H₂S, HNO₃, and O₃ in the atmosphere worldwide [9–11]. Some studies have shown that the distribution characteristics of atmospheric pollutants are closely related to meteorological elements and are affected by seasonal temperature, relative humidity, atmospheric pressure, wind direction, and wind speed. Scholars have studied the sources of various air pollutants in different research areas and consider that the sources of sulfur oxides and nitrogen oxides are industrial waste gas and biomass (coal, petroleum, etc.) combustion [12–15], while the gas conversion process of particles, road dust, and the main emissions combustion of biomass are the main sources of atmosphere pollutants [16–26].

Zhanjiang is a typical tropical ocean climate city, and the concentration levels and sources of air pollutants have not been thoroughly studied. Zhanjiang city is located at the southernmost tip of China mainland also in the southwestern of Guangdong Province and adjacent to the Tropic of Cancer in the north. Tropical and subtropical monsoon climates meet here and are regulated by the marine climate throughout the year [27]. Zhanjiang has no severe cold in winter and scorching heat in summer. According to statistics, the number of permanent residents in Zhanjiang in the past three years has been increasing yearly, at 8.30 million in 2016, 8.34 million in 2017, and 8.85 million in 2018. Moreover, Zhanjiang has vigorously developed its industry in recent years. Baosteel Zhanjiang Iron and Steel Co., Ltd. (Zhanjiang, China) officially started the construction at the Donghai Island of Zhanjiang on 31 May 2012 and was officially put into operation on 15 July 2016. In February 2019, BASF, a large German chemical company, officially announced the establishment of a new integrated base of fine chemical products on the Donghai Island of Zhanjiang. Due to the increase in the number of people and the large amount of pollutants caused by the introduction of chemical companies, the quality of atmospheric environment in Zhanjiang will undergo serious changes. At present, relatively few studies have assessed the levels and sources of air pollutants in Zhanjiang city.

Therefore, in this work, seasonal monitoring of atmospheric pollutants and total suspended particles (TSP) was carried out at four sites in coastal city, coastal urban area, coastal tourist area, and coastal industrial area. Atmospheric pollutant samples were collected by solution absorption method and analyzed by ion chromatography. The aims of this research were (1) to identify the spatiotemporal TSP variation in coastal city of Zhanjiang; (2) to explore spatiotemporal variation of atmospheric anions in coastal city Zhanjiang; and (3) to clarify the relationships among the TSP, atmospheric anions, and meteorological factors in Zhanjiang city. This study can provide a baseline scientific information for improving the coastal atmospheric environment quality in Zhanjiang city.

2. Materials and Methods

2.1. Sampling Site and Time Schedule

Four sampling sites were selected in Zhanjiang to collect atmospheric pollutants and TSP samples: the main campus of Guangdong Ocean University (S1), the Xiashan campus of Guangdong Ocean University (S2), the beach in Donghai Island (S3), and the steel base on Donghai Island (S4) (Figure 1). The latitude and longitude of each station are shown in Table 1. Among these stations, S1 is located in the coastal suburb, S2 is located in the coastal urban area, S3 is located in the coastal tourist area, and S4 is in the coastal industrial area. From June 2018 to May 2019, 8–10 days of sunny weather in the representative month of each season were selected to conduct 24-h continuous sampling at four survey stations. The 24-h continuous sampling is divided into 4 time periods (0:00–6:00 h, 6:00–12:00 h, 12:00–18:00 h, and 18:00–24:00 h) at each site for collection. A total of 228 atmospheric absorption samples and 228 TSP samples were collected.
Table 1. Longitude and latitude of each station and the distance from sea.

| Station | Longitude (N) | Latitude (E) | Distance from Sea (km) |
|---------|---------------|--------------|------------------------|
| S1      | 21°9′40.68″   | 110°17′43.44″| 6.3                    |
| S2      | 21°11′8.52″   | 110°23′35.52″| 0.9                    |
| S3      | 21°1′33.59″   | 110°31′32.52″| 0.3                    |
| S4      | 21°1′33.95″   | 110°31′13.44″| 2.0                    |

2.2. Sampling Instruments and Methods

Two parallel air TSP integrated samplers (Laoying 2050), equipped with a dust sampler and two air absorption collectors, were used to collect simultaneously coastal atmospheric pollutants and TSP samples. For air-soluble anions, each sampler was placed with two samples of 10 mL atmospheric absorption liquid (5% hydrogen peroxide); one was used to collect anions in the atmosphere, and another was used to collect blank samples at a zero flow rate. During the sampling process, one dust sampler with the GFF filter of 0.3-μm aperture was used to collect TSP samples in the atmosphere, and the other was used to collect a blank sample at a zero flow rate.

During sample collection, the flow rate of 0.2 L/min was used to collect atmospheric anion samples and zero flow rate for blank samples; the flow rate of 100 L/min was used to collect TSP samples and zero flow rate for blank TSP samples; the sampling was set to 6 h. While the sampler was working, weather data (temperature, relative humidity, pressure, wind speed, wind direction) were recorded synchronously by the PH-II-C handheld weather station. The concentration of the samples obtained in this experiment had subtracted the blank value.

2.3. Sample Pretreatment and Testing

2.3.1. Sample Pretreatment

Before sampling, 10 mL of 5% aqueous solution of hydrogen peroxide was prepared as an atmospheric absorption liquid. The 1 μm glass fiber filter membrane was first placed in a blast drying oven and baked at 105 °C for 3 h, then removed into a desiccator, and cooled at room temperature for 24 h. The electronic analytical balance (METTLER TOLEDO) was used for multiple weightings until the absolute value of the difference between the last two masses did not exceed 0.5 mg, and the average of the last two masses was recorded. Then, the fiber filter membrane was placed in the filter cartridge for TSP sampling.
2.3.2. Sample Testing

Testing of TSP Concentration

The filter with the TSP sample was weighed, and then, the mass concentration of TSP was obtained by the following formula.

\[ C = \frac{(W_1 - W_0) \times 10^3}{V_s} \]

In the formula, \( C \) represents the mass concentration of TSP (mg/m\(^3\)), \( W_0 \) represents the mass (g) of the filter before sampling, \( W_1 \) represents the mass (g) of the filter after sampling, and \( V_s \) represents the sampling volume (m\(^3\)) under the standard condition (273.15 K, 101.325 kPa).

Testing of Anion Concentration in Atmospheric Absorption Liquid

The atmospheric absorption liquid sample was ultrasonically shaken for 1 h so that the water-soluble ions adsorbed on the particulate matter were sufficiently dissolved in water. Then, the sample was filtered through a 0.45-\(\mu\)m micro-porous membrane filter and finally analyzed by ion chromatograph.

Anion chromatograph (863 compact IC plus anion, Metrohm Co. Ltd., Herisau, Switzerland) was used to determine the ion concentration of atmospheric absorption liquid samples. The mobile phase was the mixed eluent of 0.0032 mol/L Na\(_2\)CO\(_3\) and 0.001 mol/L NaHCO\(_3\). The injection volume was 10 \(\mu\)L for each injection.

2.4. Data Analysis and Statistical Methods

OriginLab software was used to draw the change curve of monitoring data. The map of sampling stations was graphed by ArcGIS and Surfer software 8. One-way analysis of variance (ANOVA) was used to examine the statistical differences in data between two or more groups. Correlation analysis between variables was determined by Pearson correlation between the environmental factors and TSP and anions using SPSS software. A probability level of 0.05 was used to determine significance.

3. Results

3.1. Seasonal Meteorological Environments Variation in Zhanjiang

In this work, the basic meteorological environment of four monitoring stations in Zhanjiang was monitored for 114 days in four seasons from June 2018 to May 2019. For comparison purposes, the average data for each season were used to represent as the meteorological traits of Zhanjiang coastal atmosphere. The temperature in Zhanjiang atmosphere ranged from 16.5 \(^\circ\)C to 39.1 \(^\circ\)C, with an average of 25.2 \(^\circ\)C. The summer temperature was the highest, with an average of 30.8 \(^\circ\)C. The mean relative humidity was 82.0\%, varying from 55.5\% to 100\%, and the average relative humidity was highest in spring at 86.1\%. The average atmospheric pressure was 1008.6 hPa, varying from 989.1 to 1019.7 hPa, and the average atmospheric pressure was lowest in summer at 998.6 hPa (Table 2). Among these factors, wind direction and wind speed were represented by a rose diagram, which showed that in the four seasons, the southeast wind and the south wind were the dominant wind directions, and the wind speed was relatively strong in spring and summer (Figure 2).
Table 2. The statistics results of meteorological factors from June 2018 to June 2019.

| Season | T(°C) | RH (%) | Air Pressure (hPa) |
|--------|-------|--------|-------------------|
| Spring | 24.9  | 86.1   | 1009.8            |
| Summer | 30.8  | 79.6   | 998.6             |
| Autumn | 24.9  | 81.4   | 1011.7            |
| Winter | 19.6  | 81.2   | 1015.2            |
| Maximum| 34.1  | 100.0  | 1019.7            |
| Minimum| 16.5  | 55.5   | 989.1             |
| Average| 25.2  | 82.0   | 1008.6            |

Figure 2. Wind rose diagram of Zhanjiang City from June 2018 to June 2019.

3.2. Seasonal Air Pollutant Concentration Variation in Zhanjiang Atmosphere

The average concentrations of coastal atmospheric pollutants of four seasons of each station are summarized in Table 3. The whole concentration levels of coastal atmospheric pollutants for four sampling stations were as follows: the coastal atmospheric chloride ion concentration ranged from 11.0 µg/m³ to 69.5 µg/m³ with the average of 29.8 µg/m³; the coastal atmospheric nitrate ion from 5.3 µg/m³ to 51.6 µg/m³ with the average of 19.6 µg/m³; the sulfate ion from 8.5 µg/m³ to 84.3 µg/m³ with the average of 45.6 µg/m³; the phosphate ion from 3.6 µg/m³ to 40.1 µg/m³ with the average of 13.5 µg/m³; and TSP from 0.03 µg/m³ to 0.59 mg/m³ and with the average of 0.28 mg/m³. There were obvious differences in the concentration of air pollutants at each survey station. The average concentrations of Cl\(^-\), NO\(_3\)\(^-\), and TSP were the highest at S2 station; the average concentration of PO\(_4\)\(^{3-}\) was the highest at S3 station; and the average concentration of SO\(_4\)\(^{2-}\) was the highest at S4 station. In summary, the pollutant concentration of each station varied greatly, and the air pollutant concentrations were also different at different stations.
With the increase of temperature, the concentration of Cl\(^-\) increased, while the concentration of PO\(_4^{3-}\) was significantly higher, and the concentrations of PO\(_4^{3-}\) were lower. With the increase of temperature, the concentration of Cl\(^-\), NO\(_3^-\), and SO\(_4^{2-}\) were significantly higher, and the concentrations of PO\(_4^{3-}\) and TSP were lower. With the increase of temperature, the concentration of Cl\(^-\), NO\(_3^-\), and SO\(_4^{2-}\) increased, while the concentration of PO\(_4^{3-}\) and TSP decreased with the increase of temperature.

### Table 3. Seasonal air pollutant concentration in each station from June 2018 to June 2019.

| Station | Range and Average | Cl\(^-\) (µg/m\(^3\)) | NO\(_3^-\) (µg/m\(^3\)) | SO\(_4^{2-}\) (µg/m\(^3\)) | PO\(_4^{3-}\) (µg/m\(^3\)) | TSP (mg/m\(^3\)) |
|---------|-------------------|--------------------------|--------------------------|--------------------------|--------------------------|------------------|
| S1      | Maximum           | 53.5                     | 29.7                     | 52.0                     | 31.2                     | 0.45             |
|         | Minimum           | 12.1                     | 5.3                      | 8.5                      | 5.0                      | 0.08             |
|         | Average           | 26.4 ± 17.6              | 19.0 ± 5.2               | 29.6 ± 14.9              | 10.6 ± 13.6              | 0.26 ± 0.17      |
| S2      | Maximum           | 69.5                     | 51.6                     | 72.2                     | 32.5                     | 0.59             |
|         | Minimum           | 12.7                     | 10.8                     | 28.6                     | 3.6                      | 0.05             |
|         | Average           | 33.0 ± 19.5              | 25.1 ± 7.0               | 50.3 ± 13.7              | 15.1 ± 11.1              | 0.34 ± 0.22      |
| S3      | Maximum           | 60.8                     | 24.8                     | 70.6                     | 29.0                     | 0.58             |
|         | Minimum           | 11.0                     | 7.6                      | 15.0                     | 8.1                      | 0.03             |
|         | Average           | 30.5 ± 16.1              | 14.7 ± 4.6               | 49.1 ± 16.2              | 17.2 ± 10.7              | 0.25 ± 0.20      |
| S4      | Maximum           | 62.1                     | 32.0                     | 84.3                     | 40.1                     | 0.41             |
|         | Minimum           | 17.2                     | 11.5                     | 10.5                     | 5.4                      | 0.12             |
|         | Average           | 29.3 ± 16.5              | 19.5 ± 3.9               | 53.2 ± 18.1              | 11.0 ± 11.0              | 0.28 ± 0.19      |
| Sum     | Maximum           | 69.5                     | 51.6                     | 84.3                     | 40.1                     | 0.59             |
|         | Minimum           | 11.0                     | 5.3                      | 8.5                      | 3.6                      | 0.03             |
|         | Average           | 29.8 ± 11.1              | 19.6 ± 4.7               | 45.6 ± 16.1              | 13.5 ± 12.4              | 0.28 ± 0.18      |

### 3.3. Coastal Atmospheric Pollutants Concentrations Variation with Meteorological Factors

#### 3.3.1. Coastal Atmospheric Pollutants Concentrations Variation Affected by Temperature

During the sampling period, the highest, lowest, and average temperatures in Zhanjiang were 34.1 °C, 16.5 °C, and 25.2 °C, respectively, and for the sake of discussion, four temperature ranges were divided as 15.0–19.9 °C, 20.0–24.9 °C, 25.0–29.9 °C, and 30.0–34.9 °C, respectively. The changes of TSP concentration and coastal atmospheric anion concentrations with temperature during the entire sampling period are illustrated in Figure 3. In the temperature range of (30–34.9) °C, the concentrations of Cl\(^-\), NO\(_3^-\), and SO\(_4^{2-}\) were significantly higher, and the concentrations of PO\(_4^{3-}\) and TSP were lower. With the increase of temperature, the concentration of Cl\(^-\), NO\(_3^-\), and SO\(_4^{2-}\) increased, while the concentration of PO\(_4^{3-}\) and TSP decreased with the increase of temperature.

![Figure 3. Relationship between temperature and atmospheric pollutant concentration.](image)
3.3.2. Coastal Atmospheric Pollutants Concentrations Variation with Relative Humidity

During the sampling period, the highest, lowest, and average relative humidity in Zhanjiang were 81.0–80.9%, 71.0–80.9%, and 81.0–90.9%, respectively. During the entire sampling period, the concentrations of TSP and coastal atmospheric anions also varied with relative humidity (Figure 4). It can be seen from the figure that when the relative humidity increased, the concentration of Cl\(^-\) and SO\(_4^{2-}\) decreased, and the concentration of NO\(_3^-\), PO\(_4^{3-}\) and TSP increased, and the concentration of NO\(_3^-\) varied very little with atmospheric pressure.

3.3.3. Coastal Atmospheric Pollutants Concentrations Variation with Atmospheric Pressure

During the sampling period, the highest, lowest, and average atmospheric pressures in Zhanjiang were 1019.7 hPa, 989.1 hPa, and 1008.6 hPa, respectively. Additionally, for the convenience of analysis, the coastal atmospheric pressure is divided into four intervals: 990.0–999.9 hPa, 1000.0–1009.9 hPa, 1010.0–1019.9 hPa, and 1020.0–1029.9 hPa. During the entire sampling period, the concentrations of TSP and coastal atmospheric anions also changed with atmospheric pressure (Figure 5). It can be seen from the figure that with the increase of atmospheric pressure, the concentration of Cl\(^-\) and SO\(_4^{2-}\) decreased, the concentration of PO\(_4^{3-}\) and TSP increased, and the concentration of NO\(_3^-\) varied very little with atmospheric pressure.

3.3.4. Coastal Atmospheric Pollutants Concentrations Variation with Wind Speed and Direction

During the sampling period, the lowest, highest, and average wind speed in Zhanjiang was 0 m/s, 6.2 m/s, and 2.7 m/s, respectively. In this study, the wind speed was divided into four intervals: 0–1.5 m/s, 1.6–3.0 m/s, 3.1–4.5 m/s, and 4.6–6.0 m/s. The TSP and PO\(_4^{3-}\) concentration decreased as the wind speed increased.
into four intervals: 0–1.5 m/s, 1.6–3.0 m/s, 3.1–4.5 m/s, and 4.6–6.0 m/s. The TSP and coastal atmospheric anion concentrations also varied with wind speed during the whole sampling period (Figure 6). As can be seen from the figure, the concentrations of each anion and TSP in the atmosphere decreased as the wind speed increased. It is noteworthy that when the wind speed was 4.6–6.0 m/s, the concentration of SO\(_4^{2-}\) dropped sharply.

![Figure 6. Relationship between wind speed and atmospheric pollutant concentration.](image)

3.3.5. Correlation Analysis of Coastal Atmospheric Anion Concentrations and Meteorological Factors

Using statistical analysis methods by SPSS software, we can quantitatively understand the relationship of coastal atmospheric anion concentration changes with various meteorological factors. The corresponding results of correlation analysis between atmospheric pollutants and meteorological factors are tabulated in Table 4. The result showed that the Cl\(^-\) concentration in the coastal atmosphere had a very significant positive correlation with air temperature and relative humidity and a negative correlation with atmospheric pressure and wind speed (Table 4). High temperature and relative humidity would promote high Cl\(^-\) concentration due to the significant positive correlations with each other. Atmospheric pressure and wind speed displayed a reverse contribution to the Cl\(^-\) concentration due to the negative correlations with each other. The correlation between NO\(_3^-\) concentration and meteorological factors was not obvious, showing that NO\(_3^-\) was less affected by meteorological factors. The concentration of SO\(_4^{2-}\) was negatively correlated with atmospheric pressure and wind speed and had no significant correlation with temperature and relative humidity, showing that SO\(_4^{2-}\) was less affected by temperature and relative humidity. The concentration of PO\(_4^{3-}\) had a negative correlation with temperature and wind speed and a positive correlation with atmospheric pressure and relative humidity, showing that the concentration of PO\(_4^{3-}\) displayed a decreasing trend with the increase of temperature and wind speed while displaying the synchronous changing trend as regards atmospheric pressure and relative humidity. The concentration of TSP displays a negative correlation with temperature, wind speed, and relative humidity but not atmosphere pressure. In summary, the concentration level and variation of atmospheric pollutants in Zhanjiang coastal atmosphere were influenced synthetically by multi-meteorological factors.
3.4. Seasonal Variation of Coastal Atmospheric Pollutants Concentration

According to the data analysis, it showed that when the season changed, the concentration of coastal atmospheric anions had obvious characteristics (Figure 7). The concentration of Cl\(^-\) and SO\(_4^{2-}\) in the atmosphere was higher in summer and lower in winter. With the change of seasons, the difference in NO\(_3^-\) concentration was not obvious. Meanwhile, the concentration of PO\(_4^{3-}\) followed the sequence of autumn > winter > spring > summer, and the concentration of TSP followed the sequence of winter > spring > autumn > summer.

| Anion      | Concentration (μg/m\(^3\)) |
|------------|----------------------------|
| Cl\(^-\)   | 0.343 **                   |
| NO\(_3^-\) | 0.027                      |
| SO\(_4^{2-}\)| 0.140                   |
| PO\(_4^{3-}\)| -0.216 *             |
| TSP        | -0.303 **                  |

| Anion concentration (μg/m\(^3\)) | Temperature (°C) | Relative Humidity (%) | Atmospheric Pressure (hPa) | Wind Speed (m·s\(^{-1}\)) |
|----------------------------------|-----------------|-----------------------|---------------------------|--------------------------|
| Cl\(^-\)                         | 0.343 **        | 0.315 **              | -0.238 *                  | -0.204 *                 |
| NO\(_3^-\)                       | 0.027           | -0.014                | -0.010                    | -0.021                   |
| SO\(_4^{2-}\)                    | 0.140           | -0.197                | -0.249 *                  | -0.266 *                 |
| PO\(_4^{3-}\)                    | -0.216 *        | 0.224 *               | 0.217 *                   | -0.202 *                 |
| TSP                              | -0.303 **       | -0.224 *              | 0.291 *                   | -0.325 **                |

* Refers to the correlation is significant at \(p < 0.05\) (two-tailed). ** Refers to the correlation is significant at \(p < 0.01\).

3.5. Changes in Atmospheric Pollutant Concentration in Different Sites

The four sampling stations represented four typical types of coastal environments in Zhanjiang atmosphere, and their distances to the sea were also different (Table 1). For the different stations, the Cl\(^-\), NO\(_3^-\), PO\(_4^{3-}\), and coastal atmospheric TSP concentrations were the highest at S2 (coastal urban area), and the SO\(_4^{2-}\) concentration at S4 (coastal industrial area) was the highest. The CI\(^-\) and SO\(_4^{2-}\) concentrations were the lowest at S1 (coastal suburb), and the NO\(_3^-\), PO\(_4^{3-}\), and TSP concentrations were the lowest at S3 (coastal tourist area). Overall, compared with other anions, the SO\(_4^{2-}\) concentration was rather higher, and the concentration of various air pollutants varied significantly with the type of environment (Figure 8).

![Figure 7](image-url) Atmospheric pollutant concentrations in different seasons.
4. Discussion

4.1. Impact of Meteorological Factors on Coastal Atmospheric Pollutants Concentrations

Temperature has a significant effect on the changes of anion concentrations and TSP concentration in the coastal atmosphere (Figure 3). It may because as the temperature increases, seawater will evaporate into the coastal atmosphere, and sea salt particles and further oxidized biogenic sulfide dimethyl sulfide (DMS) in the seawater will migrate into the atmosphere, resulting in high concentrations of Cl$^-$ and SO$_4^{2-}$ in the tropical coastal atmosphere [28]. Automobile emissions are the main source of NO$_3^-$ [29]; therefore, the NO$_3^-$ concentration does not change significantly with temperature. The concentration of PO$_4^{3-}$ and TSP decreases with increasing temperature because PO$_4^{3-}$ and TSP are mainly land-based pollutants [30]. The high-temperature weather in Zhanjiang is often probably accompanied by rainfall. Under high-temperature and rainy conditions, the PO$_4^{3-}$ and TSP in the atmosphere settle down with the rain, so the PO$_4^{3-}$ and TSP decrease as the temperature rises. Zhanjiang city is located in a subtropical region and belongs to a subtropical monsoon climate with sufficient rainfall. Meanwhile, the ocean atmosphere along with the subtropical monsoon blows to the mainland, which brings a large amount of water vapor to ensure a high relative humidity in Zhanjiang atmosphere. With the increase of relative humidity, the land-derived pollutants NO$_3^-$, PO$_4^{3-}$, and TSP will partly settle with the condensation of water vapor, resulting in the decrease of the concentration of NO$_3^-$, PO$_4^{3-}$, and TSP in the atmosphere (Figure 4). The ocean atmosphere contains a large amount of water vapor, sea salt particles, and sulfate ions, which increase the concentration of Cl$^-$ and SO$_4^{2-}$ in the coastal atmosphere [31]. In addition, when the atmospheric pressure increases (Figure 5), the air convection is not strong, and the rainfall is less, so the path of the Cl$^-$ and SO$_4^{2-}$ transported through the ocean atmosphere is blocked, resulting in an overall decrease in the concentration of Cl$^-$ and SO$_4^{2-}$, and with the decrease of tropical high-pressure rainfall, the accumulation of PO$_4^{3-}$ and TSP in the atmosphere has an increasing trend. Regarding wind speed, studies showed that strong wind could help to remove most of air pollutants in the coastal atmosphere [32]. Combined with the analysis of air pollutant concentration and wind speed in Zhanjiang (Figure 6), it shows that when the wind speed increases, the concentration of air pollutants and TSP both are reduced, which is consistent with the above research results.

4.2. Effects of Different Seasons on Coastal Atmospheric Pollutants Concentrations

Seasonal changes will affect the concentration of coastal atmosphere pollutants. The concentration of Cl$^-$ and SO$_4^{2-}$ is higher in summer because of the dominant southeast wind and high temperature of Zhanjiang atmosphere in summer. A large number of sea
salt particles evaporates from the surface seawater into the coastal atmosphere, and a large amount of SO$_4^{2-}$ and Cl$^-$ are attached to the sea salt particles and are continuously transported to the terrestrial atmosphere with the southeast wind [33–36]. Moreover, the metabolism of seaweed and plankton metabolism also had a certain contribution to the SO$_4^{2-}$ concentration [37,38]. In Zhanjiang’s autumn and winter seasons, the dominant wind direction is northwest wind. Although a small amount of Cl$^-$ and SO$_4^{2-}$ comes from exhaust emissions from factories and automobiles [39], Cl$^-$ and SO$_4^{2-}$ mainly come from the ocean. Therefore, the concentration of Cl$^-$ and SO$_4^{2-}$ in autumn and winter is lower than in summer due to a lower temperature. In addition, the concentration of NO$_3^-$ fluctuates slightly: the concentration in winter is slightly higher, and the concentration in summer is slightly lower. It may be that NO$_3^-$ mainly comes from the combustion of fuels in automobiles [40–42], factories [43] and other industries [44] as well as the combustion of trees, crops, and other plants [45,46]. These land-derived pollutants are blown from inland to the coast under the prevailing northwest wind in winter. Therefore, in the coastal atmosphere of Zhanjiang, the concentration of NO$_3^-$ in winter is higher than in other seasons. The concentration of PO$_4^{3-}$ is mainly affected by human activities, including the industrial production of various phosphate products, poultry excrement [47], solid waste [48], the use of phosphate fertilizers [49], and coal combustion [50]. The concentration of PO$_4^{3-}$ in winter and winter is higher than that in spring and summer, mainly because there is more rainfall in spring and summer, which will make some of the PO$_4^{3-}$ in the atmosphere diminish. The concentration of TSP in the atmosphere is affected by natural and human factors. The natural influence is mainly from the accumulation of atmospheric particulate matter caused by wind and dust in the west and north direction [51]. Compared with natural sources of impact, human sources of impact are more complex, including industrial boilers [52], waste incineration [53,54], household heating [55], construction [56], and mining [57,58]. In general, TSP is mainly a land-based pollutant. The concentration of TSP in winter is higher than that in summer, mainly because the northwest wind prevails in winter, and there is more rainfall in summer. Notably, the atmospheric TSP concentration was slightly larger in spring than in autumn due to the increase of dewdrops in spring.

4.3. Impact of Different Environmental Types on Coastal Atmospheric Pollutants Concentrations

Based on the environmental characteristics of Zhanjiang coastal atmosphere, the S1 station (the main campus of Guangdong Ocean University) belongs to the typical coastal rural environment, so the concentrations of various anions and TSP are not high in this rural area [59]. Meanwhile, Cl$^-$, SO$_4^{2-}$, and PO$_4^{3-}$ are the lowest among all investigated environmental types. Far from the sea, fewer types of enterprises and the small amount of motor vehicles are the main causes for the less polluted environment in this area. The S2 station (the Xiashan campus of Guangdong Ocean University) is a typical coastal urban environment, where the concentration level of anions and TSP in the atmosphere is the highest among all stations. Closer to the sea, a large number of motor vehicles and many urban enterprises (especially the catering industry) make a common contribution to the serious atmosphere pollution [60]. It means that fuel combustion, motor vehicle emissions, and the contributions from the ocean are the main sources of air pollution in the coastal urban atmosphere of Zhanjiang. The S3 station (the beach in Donghai Island), which is close to the area of steel industry, is a typical coastal atmospheric environment, so the pollutant concentration in this area was generally maintained at a medium level, and the main pollution cause was the contribution of evaporating from seawater and the diffusion effect of pollutants from the nearby steel industry atmosphere [61]. The S4 station (the steel base in Donghai Island) is a typical coastal industrial environment, so the concentration of various polluting anions and TSP in this area is relatively high. Its main sources of NO$_x$ and SO$_x$ are the contribution of evaporation from seawater and the industrial atmospheric emissions. However, due to the non-continuous and limited data at present, the different source contributions should be traced and explored by on-line monitoring and dynamic
models in depth, which is critical to protect the Zhanjiang air quality under the great anthropogenic emission pressure in the future.

5. Conclusions

There are various kinds of air pollutants that have an impact on air environmental quality. Soluble anions and total suspended particles are two important types of air pollutants. In this work, through the on-site monitoring in four typical environmental areas of Zhanjiang city atmosphere, the temporal and spatial distribution characteristics of water-soluble anions and TSP were obtained. The results indicated that the concentration changes of coastal atmospheric pollutants showed significant relationships with meteorological factors. The concentration of $\text{PO}_4^{3-}$ and TSP was negatively correlated with temperature, relative humidity, and air pressure. However, the concentrations of $\text{Cl}^-$ and $\text{SO}_4^{2-}$ were positively correlated with those environmental factors. In addition, as the wind speed increased, the concentration of each pollutant decreased as a whole. Furthermore, sea wind helps to raise the concentration of soluble anions in atmosphere. Moreover, the concentrations of various coastal anions and TSP changed greatly in different seasons. During sampling periods, the concentrations of $\text{Cl}^-$ and $\text{SO}_4^{2-}$ were largest in summer and lowest in winter. The differences in $\text{NO}_3^-$ concentration were not significant between seasonal changes. Comparatively, the concentrations of anions and TSP follow the order of coastal urban atmosphere $>$ coastal industrial area $\approx$ coastal tourist area $>$ coastal suburb in Zhanjiang coastal atmosphere. The contributions of motor vehicle emissions, urban enterprise emissions, and seawater evaporation were the main causes of air pollution in Zhanjiang coastal atmosphere. To avoid the deterioration decline of air quality, the mitigation strategies should be traced to the emission sources of air pollutants in the future.

Author Contributions: Conceptualization, J.-B.Z. and P.Z.; methodology, J.-B.Z., Y.-M.R., Q.-F.Y. and P.Z.; software, P.Z.; validation, J.-B.Z., Y.-M.R., Q.-F.Y. and P.Z.; formal analysis, Y.-M.R. and Q.-F.Y.; investigation, J.-B.Z., Y.-M.R. and Q.-F.Y.; resources, C.-L.C.; data curation, L.-R.Z.; writing—original draft preparation, J.-B.Z. and Q.-F.Y.; writing—review and editing, J.-B.Z., Y.-M.R., Q.-F.Y., and P.Z.; visualization, J.-B.Z. and L.-R.Z.; supervision, J.-B.Z. and C.-L.C.; project administration, J.-B.Z. and C.-L.C.; funding acquisition, J.-B.Z. All authors have read and agreed to the published version of the manuscript.

Funding: This research was funded by Science and Technology Program Project of Guangdong Province: Corrosion Failure Prediction of Coastal Amusement Facilities in Tropical Marine Environment, grant number 2016A020225004, and the Research and Development Projects in Key Areas of Guangdong Province (2020B1111020004), and Innovation Strong School Project of Guangdong Ocean University (230420021).

Institutional Review Board Statement: Not applicable.

Informed Consent Statement: Not applicable.

Data Availability Statement: Data sharing is not applicable to this article.

Acknowledgments: The authors are grateful for the anonymous reviewers’ careful review and constructive suggestions to improve the manuscript. Thanks are given to Guangdong Province Science and Technology Program Project Fund (2016A020225004) and the Research and Development Projects in Key Areas of Guangdong Province (2020B1111020004).

Conflicts of Interest: The authors declare no conflict of interest.

References

1. Xu, X.M.; Zhang, H.F. Six sources mainly contributing to the haze episodes and health risk assessment of PM$_{2.5}$ at Beijing suburb in winter 2016. *Ecotox. Environ. Saf.* 2018, 166, 146–156. [CrossRef] [PubMed]
2. Andini, A.; Bonnet, S.; Rousset, P.; Hasanudin, U. Impact of open burning of crop residues on air pollution and climate change in Indonesia. *Curr. Sci.* 2018, 115, 2259–2266. [CrossRef]
3. Buszman, P.E.; Derbisz, K.; Kwasiborski, P.; Chrząszcz, P.; Mularska, M.; Baron, D.; Sobieszek, A.; Mendyk, A.; Skoczylas, P.; Cisowski, M.; et al. Impact of air pollution on hospital patients admitted with ST- and non-ST-segment elevation myocardial infarction in heavily polluted cities within the European Union. *Cardiol. J.* 2020, 27, 541–547. [CrossRef] [PubMed]

4. Chen, C.; Li, C.C.; Li, Y.W.; Liu, J.Y.; Meng, C.S.; Han, J.X.; Zhang, Y.W.; Xu, D.Q. Short-term effects of ambient air pollution exposure on lung function: A longitudinal study among healthy primary school children in China. *Sci. Total Environ.* 2018, 645, 1014–1020. [CrossRef] [PubMed]

5. Duan, J.; Cheng, Q.; Luo, X.L.; Bai, L.J.; Zhang, H.; Wang, S.S.; Xu, Z.H.; Gao, J.J.; Zhang, Y.W.; Su, H. Is the serious ambient air pollution associated with increased admissions for schizophrenia? *Sci. Total Environ.* 2018, 644, 14–19. [CrossRef] [PubMed]

6. Lin, S.Y.; Hsu, W.H.; Lin, C.L.; Lin, C.C.; Lin, C.H.; Hsu, I.K.; Hsu, C.Y.; Kao, C.H. Association of Exposure to Fine-Particulate Air Pollution and Acidity Gases with Incidence of Nephrotic Syndrome. *Int. J. Environ. Res. Public Health* 2018, 15, 2860. [CrossRef]

7. Rossi, G.; Vigotti, M.A.; Zanobetti, A.; Repetto, F.; Gianelle, V.; Schwartz, J. Air pollution and cause-specific mortality in Milan, Italy, 1980–1989. *Arch. Environ. Health Int. J.* 1999, 54, 158–164. [CrossRef]

8. Ware, J.H.; Ferris, B.G.; Dockery, D.W.; Spengler, J.D.; Straum, D.O.; Speizer, F.E. Effects of ambient sulfur oxides and suspended particles on respiratory health of preadolescent children. *Am. Rev. Respir. Dis.* 1986, 135, 834–842.

9. Hong, Q.Q.; Liu, C.; Hu, Q.H.; Xing, C.Z.; Tan, W.; Liu, H.R.; Huang, Y.; Zhu, Y.; Zhang, J.S.; Geng, T.Z.; et al. Evolution of the vertical structure of air pollutants during winter heavy pollution episodes: The role of regional transport and potential sources. *Atmo. Res.* 2019, 228, 206–222. [CrossRef]

10. Liu, Z.; Qi, J.H.; Wang, L.; Chen, X.J.; Shi, J.H.; Gao, H.W. Seasonal distribution of water-soluble inorganic ions in the atmospheric aerosol in Qingdao. *Huanjing Kexue* 2012, 33, 2180–2190.

11. Manju, A.; Kalaiselvi, K.; Dhananjanay, V.; Palanivel, M.; Banupriya, G.S.; Vishy, M.H.; Panjakumar, K.; Ravichandran, B. Spatio-seasonal variation in ambient air pollutants and influence of meteorological factors in Coimbatore, Southern India. *Air Qual. Atmos. Health* 2018, 11, 1179–1189. [CrossRef]

12. Zhao, H.J.; Ma, Y.J.; Wang, Y.Q. Aerosol and gaseous pollutant characteristics during the heating season (winter-spring transition) in the Harbin-Changchun megalopolis, northeastern China. *J. Atmos. Sol.-Terr. Phys.* 2019, 188, 26–43. [CrossRef]

13. Zhang, J.J.; Tong, L.; Huang, Z.W.; Zhang, H.L.; He, M.M.; Dai, X.R.; Zheng, J.; Xiao, H. Seasonal variation and size distributions of water-soluble inorganic ions and carbonaceous aerosols at a coastal site in Ningbo, China. *Sci. Total Environ.* 2018, 639, 793–803. [CrossRef] [PubMed]

14. Xu, J.S.; Xu, M.X.; Snape, C.; He, J.; Behera, S.N.; Xu, H.H.; Ji, D.S.; Wang, C.J.; Yu, H.; Xiao, H.; et al. Temporal and spatial variation in major ion chemistry and source identification of secondary inorganic aerosols in Northern Zhejiang Province, China. *Chemosphere* 2017, 179, 316–330. [CrossRef] [PubMed]

15. Li, Y.P.; Hao, Q.J.; Wen, T.X.; Ji, D.S.; Liu, Z.R.; Wang, Y.S.; Jiang, C.S. Mass Concentrations and Size Distributions of Water-soluble Inorganic Ions in Atmospheric Aerosols in Beibei District, Chongqing. *Huanjing Kexue* 2018, 39, 4002–4013. [PubMed]

16. Cesari, D.; Genga, A.; Ielpo, P.; Siciliano, M.; Mascavo, G.; Grasso, F.M.; Contini, D. Source apportionment of PM2.5 in the harbour-industry area of Brindisi (Italy): Identification and estimation of the contribution of in-port ship emissions. *Sci. Total Environ.* 2014, 497, 392–400. [CrossRef] [PubMed]

17. Arideep, M.; Madhoolika, A. Assessment of local and distant sources of urban PM2.5, in middle Indo-Gangetic plain of India using statistical modeling. *Atmos. Res.* 2018, 213, 275–287.

18. Bozkurt, Z.; Uzmez, O.O.; Dogeroglu, T.; Artun, G.; Gaga, E.O. Atmospheric concentrations of SO2, NO2, ozone and VOCs in Duzce, Turkey using passive air samplers: Sources, spatial and seasonal variations and health risk estimation. *Atmos. Res.* 2018, 9, 1146–1156. [CrossRef] [PubMed]

19. Jeričević, A.; Gašparac, G.; Maslač, M.M.; Kumar, P.; Prtenjak, M.T. Identification of diverse air pollution sources in a complex urban area of Croatia. *J. Environ. Manag.* 2019, 243, 67–77. [CrossRef] [PubMed]

20. Dimitriou, K.; Kassomenos, P. A meteorological analysis of PM10 episodes at a high altitude city and a low altitude city in central Greece—The impact of wood burning heating devices. *Atmos. Res.* 2018, 214, 329–337. [CrossRef] [PubMed]

21. Beddows, D.C.S.; Harrison, R.M. Identification of specific sources of airborne particles emitted from within a complex industrial (steelworks) site. *Atmos. Environ.* 2018, 183, 123–134. [CrossRef]

22. Schade, G.W.; Roest, G. Source apportionment of non-methane hydrocarbons, NOx and H2S data from a central monitoring station in the Eagle Ford shale, Texas. *Elementa-Sci. Anthrop.* 2018, 6, 35. [CrossRef]

23. Perrone, M.G.; Vratolis, S.; Georgieva, E.; Torok, S.; Sega, K.; Veleba, B.; Osan, J.; Beslic, I.; Kertesz, Z.; Pernigotti, D.; et al. Sources and geographic origin of particulate matter in urban areas of the Danube macro-region: The cases of Zagreb (Croatia), Budapest (Hungary) and Sofia (Bulgaria). *Sci. Total Environ.* 2018, 619, 1515–1529. [CrossRef] [PubMed]

24. Li, T.C.; Chen, W.H.; Yuan, C.S.; Wu, S.P.; Wang, X.H. Physicochemical Characteristics and Source Apportionment of Atmospheric Aerosol Particles in Kinmen-Xiamen Airshed. *Environ. Eng.* 2013, 13, 308–323. [CrossRef]

25. Tian, M.; Wang, H.; Chen, Y.; Yang, F.M.; Zhang, X.H.; Zou, Q.; Zhang, R.Q.; Ma, Y.L.; He, K.B. Characteristics of aerosol pollution during heavy haze events in Suzhou, China. *Atmos. Chem. Phys.* 2016, 16, 7357–7371. [CrossRef]

26. Bahino, J.; Yoboue, V.; Galy-Lacaux, C.; Adon, M.; Akpo, A.; Keita, S.; Liousse, C.; Gardrat, E.; Chiron, C.; Ossohou, M.; et al. A pilot study of gaseous pollutants’ measurement (NO2, SO2, NH3, HNO3 and O3) in Abidjan, Cote d’Ivoire: contribution to an overview of gaseous pollution in African cities. *Atmos. Chem. Phys.* 2018, 243, 67–77.
27. Zhang, J.B.; Yang, B.; Chen, T.; Sun, X.L.; Chen, C.L. Metal speciation and pollution assessment of Cd and Pb in intertidal sediments of Donghai Island, China. Reg. Stud. Mar. Sci. 2016, 6, 37–48.

28. George, S.K.; Nair, P.R.; Parameswaran, K.; Jacob, S.; Abraham, A. Seasonal trends in chemical composition of aerosols at a tropical coastal site of India. J. Geophys. Res.-Atmos. 2008, 113, D16209. [CrossRef]

29. Branis, M. Air quality of Prague: Traffic as a main pollution source. Environ. Monit. Assess. 2009, 156, 377–390. [CrossRef]

30. Zehra, B. Seasonal variation of water-soluble inorganic ions in PM\(_{2.5}\) in a city of northwestern Turkey. Environ. Forensics 2018, 19, 1–13.

31. Cataldo, J.; González, A.E. Analysis of the Relation between Particle Matter Concentration and Meteorological Parameter at Montevideo City. Sci. Res. Pub. 2018, 6, 120–139. [CrossRef]

32. Xie, J.L.; Liao, Z.H.; Fang, X.Q.; Xu, X.Q.; Wang, Y.; Zhang, Y.; Liu, J.; Fan, S.J.; Wang, B.M. The characteristics of hourly wind field and its impacts on air quality in the Pearl River Delta region during 2013–2017. Atmos. Res. 2019, 227, 112–124. [CrossRef]

33. Jusino-Aresino, R.; Anderson, J.; Gao, Y. Ionic and elemental composition of PM\(_{2.5}\) aerosols over the Caribbean Sea in the Tropical Atlantic. J. Atmos. Chem. 2016, 73, 427–457. [CrossRef]

34. Tang, Y.S.; Braban, C.F.; Dragosits, U.; Simmonds, I.; Leaver, D.; van Dijk, N.; Poskitt, J.; Thacker, S.; Patel, M.; Carter, H.; et al. Acid gases and aerosol measurements in the UK (1999–2015): Regional distributions and trends. Atmos. Chem. Phys. 2018, 18, 16293–16324. [CrossRef]

35. Wang, Q.D.; Song, J.M.; Li, X.G.; Yuan, H.M.; Li, N.; Duan, L.Q.; Ren, C.Z. Geochemical characteristics and potential biogeochemical effect of water-soluble ions in atmospheric aerosols over the western boundary regions of Pacific Ocean. Atmos. Res. 2019, 227, 101–111. [CrossRef]

36. Wu, R.; Zhou, X.; Wang, L.P.; Wang, Z.; Zhou, Y.; Zhang, J.Z.; Wang, W.X. PM\(_{2.5}\) Characteristics in Qingdao and across Coastal Cities in China. Atmosphere 2017, 8, 77. [CrossRef]

37. Abbatt, J.P.D.; Leaitch, W.R.; Alibabdi, A.A. Overview paper: New insights into aerosol and climate in the Arctic. Atmos. Chem. Phys. 2019, 19, 2527–2560. [CrossRef]

38. Enami, S.; Sakamoto, Y.; Haru, K.; Osada, K.; Hoffmann, M.R.; Colussi, A.J. “Sizing” Heterogeneous Chemistry in the Conversion of Gaseous Dimethyl Sulphide to Atmospheric Particulates. Environ. Sci. Technol. 2016, 50, 1834–1843. [CrossRef]

39. Yang, Y.; Li, X.R.; Chen, X.; Liu, S.Q.; Liu, Y.S.; Xu, J.; Wang, L.L.; Tao, M.H.; Wang, G.H. Analysis of Different Particle Sizes, Pollution Characteristics, and Sources of Atmospheric Aerosols During the Spring Dust Period in Beijing. Environ. Sci. 2018, 39, 5315–5322.

40. Adame, J.A.; Lopez, L.; Hidalgo, P.J.; Sorribas, M.; Gutierrez-Alvarez, I.; del Aguila, A.; Saiz-Lopez, A.; Yela, M. Study of the exceptional meteorological conditions, trace gases and particulate matter measured during the 2017 forest fire in Donana Natural Park, Spain. Sci. Total Environ. 2018, 645, 710–720. [CrossRef][PubMed]

41. Brennun-Hansen, H.; Bender, A.M.; Andersen, Z.J.; Sorensen, J.; Bonlloekte, J.H.; Boshuizen, H.; Becker, T.; Diderichsen, F.; Loft, S. Assessment of impact of traffic-related air pollution on morbidity and mortality in Copenhagen Municipality and the health gain of reduced exposure. Environ. Int. 2018, 121, 973–980. [CrossRef][PubMed]

42. Salvador, P.; Molero, F.; Fernandez, A.J.; Tobias, A.; Pandolfi, M.; Gomez-Moreno, F.J.; Barreiro, M.; Perez, N. Synergistic effect of the occurrence of African dust outbreaks on atmospheric pollutant levels in the Madrid metropolitan area. Atmos. Res. 2019, 226, 208–218. [CrossRef]

43. Chen, W.; Li, A.J.; Zhang, F. Roadside atmospheric pollution: Still a serious environmental problem in Beijing, China. Air Qual. Atmos. Health 2018, 11, 1203–1216. [CrossRef][PubMed]

44. Plocoste, T.; Dorville, J.F.; Monjoly, S.; Jacoby-Koaly, S.; Andre, M. Assessment of nitrogen oxides and ground-level ozone behavior in a dense air quality station network: Case study in the Lesser Antilles Arc. J. Air Waste Manag. Assoc. 2018, 68, 1278–1300. [CrossRef][PubMed]

45. Abdul, H.N.D.; Latif, M.T.; Ahamad, F.; Dominick, D.; Chung, J.X.; Juneng, L.; Khan, M.F. The long-term assessment of air quality on an island in Malaysia. Heliyon 2018, 4, e01054. [CrossRef][PubMed]

46. Adeniran, J.A.; Yusuf, R.O.; Fakirin, B.S.; Sonibare, J.A. Air quality assessment and modelling of pollutants emission from a major cement plant complex in Nigeria. Atmos. Pollut. Res. 2019, 10, 257–266. [CrossRef]

47. Querol, X.; Alastuey, A.; de la Rosa, J.; Sanchez-de-la-Campa, A.; Plana, F.; Ruiz, C.R. Source apportionment analysis of atmospheric particulates in an industrialised urban site in southwestern Spain. Atmos. Environ. 2002, 36, 3113–3125. [CrossRef]

48. Huston, R.; Chan, Y.C.; Gardner, T.; Shaw, G.; Chapman, H. Characterisation of atmospheric deposition as a source of contaminants in urban rainwater tanks. Water Res. 2009, 43, 1630–1640. [CrossRef]

49. Karar, K.; Gupta, A.K. Seasonal variations and chemical characterization of ambient PM\(_{10}\) at residential and industrial sites of an urban region of Kolkata (Calcutta), India. Atmos. Res. 2006, 81, 36–53. [CrossRef]

50. Wang, N.; Ma, M.; Shi, J.H.; Gao, H.W.; Yao, X.H. Concentrations, Sources, and Dry Deposition Fluxes of Different Forms of Phosphorus in Qingdao Aerosols in Summer. Huanjing Kexue 2018, 39, 4034–4041.

51. Qi, J.H.; Liu, X.H.; Yao, X.H.; Zhang, R.F.; Chen, X.J.; Lin, X.H.; Gao, H.W.; Liu, R.H. The concentration, source and deposition flux of ammonium and nitrate in atmospheric particles during dust events at a coastal site in northern China. Atmos. Chem. Phys. 2018, 18, 571–586. [CrossRef]

52. Rojano, R.E.; Arregoces, H.A.; Angulo, L.C.; Restrepo, G.M. Analysis and Origin of PM\(_{10}\) and PM 10 Concentrations in Open-pit Coal Mining using Polar Plots. Inf. Technol. 2018, 29, 131–142.
53. Duan, W.J.; Liang, J.L.; Cheng, S.Y.; Jia, J.; Wang, X.Q. Air Pollutant Emission Inventory from Iron and Steel Industry in the Beijing-Tianjin-Hebei Region and Its Impact on PM(2.5). *Environ. Sci.* 2018, 39, 1445–1454.

54. Yan, H.; Ding, G.L.; Li, H.Y.; Wang, Y.S.; Zhang, L.; Shen, Q.P.; Feng, K.L. Field Evaluation of the Dust Impacts from Construction Sites on Surrounding Areas: A City Case Study in China. *Sustainability*. 2019, 11, 1906. [CrossRef]

55. Huang, Y.; Du, W.; Chen, Y.C.; Shen, G.F.; Su, S.; Lin, N.; Shen, H.Z.; Zhu, D.; Yuan, C.Y.; Duan, Y.H.; et al. Household air pollution and personal inhalation exposure to particles (TSP/PM$_{2.5}$/PM$_{1.0}$/PM$_{0.25}$) in rural Shanxi, North China. *Environ. Pollut.* 2017, 231, 635–643. [CrossRef] [PubMed]

56. Adeniran, J.A.; Yusuf, R.O.; Olajire, A.A. Exposure to coarse and fine particulate matter at and around major intra-urban traffic intersections of Ilorin metropolis, Nigeria. *Atmos. Environ.* 2017, 166, 383–392. [CrossRef]

57. Guo, Y.Y.; Gao, X.; Zhu, T.Y.; Luo, L.; Zheng, Y. Chemical profiles of PM emitted from the iron and steel industry in northern China. *Atmos. Environ.* 2017, 150, 187–197. [CrossRef]

58. Liu, T.N.; Zhuang, G.S.; Huang, K.; Lin, J.; Wang, Q.Z.; Deng, C.R.; Fu, J.S. A Typical Formation Mechanism of Heavy Haze-Fog Induced by Coal Combustion in an Inland City in North-Western China. *Aerosol Air Qual. Res.* 2017, 17, 98–107. [CrossRef]

59. Bilal, M.; Nichol, J.E.; Nazeer, M.; Shi, Y.; Wang, L.C.; Kumar, K.R.; Ho, H.C.; Mazhar, U. Characteristics of Fine Particulate Matter (PM$_{2.5}$) over Urban, Suburban, and Rural Areas of Hong Kong. *Atmosphere* 2019, 10, 496. [CrossRef]

60. Lu, H.Y.; Wu, Y.L.; Mutuku, J.K.; Chang, K.H. Various Sources of PM$_{2.5}$ and their Impact on the Air Quality in Tainan City, Taiwan. *Aerosol Air Qual. Res.* 2019, 19, 601–619. [CrossRef]

61. Sun, W.Q.; Zhou, Y.; Lv, J.X.; Wu, J.Z. Assessment of multi-air emissions: Case of particulate matter (dust), SO$_2$, NOx and CO$_2$ from iron and steel industry of China. *J. Clean Prod.* 2019, 232, 350–358. [CrossRef]