Invasion genetics: Lessons from a ubiquitous bird, the house sparrow *Passer domesticus*

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**Abstract** Following an introduction, non-native species are exposed to environments that differ from those found in their native range; further, as these non-native species expand beyond the site of introduction, they must constantly adapt to novel environments. Although introduced species are present across most ecosystems, few species have successfully established themselves on a truly global scale. One such species, the house sparrow *Passer domesticus*, is now one of the world’s most broadly distributed vertebrate species and has been introduced to a great part of its current range. To date, work on four continents suggests both genetic and phenotypic variation exists between native and introduced ranges. As such, house sparrows represent an excellent opportunity to study adaptations to novel environments and how these adaptations are derived. The global distribution of this species and the multiple independent introductions to geographically isolated sites allow researchers to ask questions regarding genetic variation and adaptation on a global scale. Here, we summarize the molecular studies of invasive house sparrows from the earliest work using allozymes through more recent work on epigenetics; using these studies, we discuss patterns of dispersal of this species. We then discuss future directions in techniques (e.g. next generation sequencing) and how they will provide new insight into questions that are fundamental to invasion biology. Finally, we discuss how continued research on the house sparrow in light of these genetic changes and adaptations will elucidate answers of adaptation, invasion biology, range expansion, and resilience in vertebrate systems generally [Current Zoology 61 (3): 465–476, 2015].

**Keywords** House sparrow, Invasive, *Passer domesticus*

1 Introduction

Introduced species are the second largest threat to global biodiversity. In the face of ecosystem impacts (Bakker and Wilson, 2004), economic costs (Mack et al., 2000), and the likelihood of growing commerce increasing the threat of new introductions (Levine and D’antonio, 2003), invasive species research has grown recently. However, in addition to invasive species management, studying invasive species in an ecological or evolutionary context will lend great insight to fundamental questions in biology such as the general spatial structure of species interactions, allopatric speciation, and response to environmental stressors and challenges (Holt, 2003).

Following an introduction, organisms cannot only endure novel environments but must also thrive in them to be successful. Additionally, as introduced species expand beyond the site of initial introduction, novel environments will likely continue to be encountered so that individuals must continuously adjust phenotypes to cope. Most introductions involve bottlenecks and founder effects that can reduce genetic diversity initially. Unlike many threatened species facing the same challenge, introduced species are often able to overcome these reductions in genetic diversity to generate novel adaptations and adjust to novel environments; this presents a genetic paradox (Dlugosch and Parker, 2008).

In novel environments, phenotypes should match new environments, thus creating phenotypic gradients consistent with age of population (Sol et al., 2002, Rehage and Sih, 2004, Russell et al., 2010). However, the underlying mechanisms of this variation in phenotypes, particularly on such short timescales, are less clear. These phenotypic changes may be due to changes in the frequency of genes across space and time, i.e. driven by natural selection, genetic drift, founder effects (Clegg et al., 2002), or spatial sorting of genes as the population expands (Shine et al., 2011). Alternatively, phenotypic variation across ranges may be determined by other fac-
tors such as epigenetic changes (Richards et al., 2012, Schrey et al., 2012, Liebl et al., 2013a) and phenotypic plasticity or flexibility in response to variable environments (Sol et al., 2002, Yeh and Price, 2004, Richards et al., 2006, Davidson et al., 2011). Epigenetic changes such as methylation induce changes in gene expression, which induce changes in phenotypes (Richards, 2006). Phenotypic plasticity and flexibility, which may be influenced by epigenetic and/or genetic factors, allow individuals to rapidly adjust to environmental variation (Agrawal, 2001). Factors other than genetic-based evolution are likely to contribute to phenotypic variation in introduced species given the likelihood of significant reductions in genetic diversity due to bottlenecks and founders effects at the time of initial introduction and throughout range expansion (Dlugosch and Parker, 2008).

Despite considerable challenges in successfully colonizing novel habitats following an introduction, some species have been repeatedly and successfully introduced throughout the world. In part, this success will reflect the amount of effort invested in introducing these species (Cassey et al., 2004, Blackburn et al., 2011). However, it is also recognized that some species have inherent characteristics that enable them to succeed in new environments, i.e. “weedy” species (Cassey et al., 2004). One such species, the house sparrow Passer domesticus, native to Europe and Asia, has been introduced and become established on four continents (Fig. 1), making it one of the most broadly distributed vertebrate species in the world (Anderson, 2006). House sparrows have been introduced intentionally (e.g. as a familiar species or pest control) or unintentionally (Anderson, 2006), however, unlike many other species, the house sparrow has successfully expanded its range in most introduced environments (Robbins, 1973; Anderson, 2006), indicating an incredible ability to overcome genetic bottlenecks and adjust to variable climates and environments.

House sparrows were introduced to North America in the 1850s and 60s from Europe and rapidly expanded its distribution to cover all of the United States and most of Canada and Mexico by the early 1900s (Grinnell, 1919). In South America, house sparrows have been introduced multiple times including to Argentina in the late 1800s, and Chile and Brazil in the early 20th century; since, house sparrows have spread to occupy most areas of South America including Tierra del Fuego (Lever, 2005). In Australia and New Zealand, house sparrows were introduced in the early 1860s (Long, 1981), currently occupying all of New Zealand and the eastern half of Australia (Higgins and Peter, 2002). Finally, house sparrows were introduced to South Africa around 1900, from where they expanded in a northerly direction reaching southern Tanzania in the early 1980s (Anderson, 2006); additionally, a secondary introduction from South Africa into the ports of Mombasa, Kenya and Dar Salam, Tanzania occurred in the 1950s (Anderson, 2006). Another introduction (of unknown origin) brought house sparrows to Senegal and western Africa in the late 1970s (Anderson, 2006) (Fig. 1).

![Fig. 1 Map of current distribution of house sparrows](https://academic.oup.com/cz/article-abstract/61/3/465/178647/fig1)

Black area indicates native range; dark gray indicates long-established introduced (> 100 years since introduction); light gray indicates recently introduced (< 100 years).
Given its broad range and the commensal nature of the species, house sparrows have been studied extensively as a model for general ecological, behavioral, and evolutionary questions (Anderson, 2006). For example, house sparrows have been used to study many phenotypic traits such as morphology (Johnston and Selander, 1964; Johnston and Selander, 1971; Johnston and Selander, 1973), personality (Martin and Fitzgerald, 2005; Lendvai et al., 2011; Bókony et al., 2012; Bókony et al., 2014), stress physiology (Lendvai and Chastel, 2010; Koren et al., 2012; Liebl et al., 2013b; Lattin and Romero, 2014), and sexual selection and reproduction (Griffith et al., 1999a; Nakagawa et al., 2007; Partecke and Schwabl, 2008; Lendvai and Chastel, 2010; Ouyang et al., 2011). The extensive research conducted in this species provides a good opportunity to draw comparisons between different populations (e.g. Griffith et al., 1999b), or different outcomes (e.g. Nakagawa et al., 2007). Such intra-specific comparisons are more powerful than inter-specific comparisons because inter-population variation is composed of ecological, rather than phylogenetic variation.

Although most of the studies conducted on house sparrows have not specifically addressed differences according to invasion status, considerable work exists on house sparrow biology (reviewed in Anderson, 2006), which means that we know a great deal about this species, which allows researchers concerned with invasiveness to better design and interpret research in this area. David Lack first identified house sparrows as a good model for studies of evolutionary biology recognizing that introduced house sparrows in North America provided an excellent opportunity to understand how adaptive processes could work over a relatively short timeframe (Lack, 1940). In fact, the work by Johnston and Selander (1964; 1967) on range expansion and morphological variation in American house sparrows has become one of the text book examples of selection and adaptation; these studies showed rapid morphological differentiation among house sparrow populations over a period of ~100 years, which challenged previous predictions on estimated population differentiation (Johnston and Selander, 1964; Selander and Johnston, 1967). This work has been followed by studies of other aspects of adaptiveness and invasiveness in relation to traits such as response to novelty (Martin and Fitzgerald, 2005; Liebl and Martin, 2012; Liebl and Martin, 2014), stress physiology (Liebl and Martin, 2012; Liebl and Martin, 2013), immune regulation (Martin et al., 2010; Martin et al., 2014; Martin et al., in press) and dispersal (Skjelseth et al., 2007). Additionally, researchers can use the large research base which exists on house sparrows for single-species meta-analyses (e.g. Nakagawa et al., 2007), which are often stronger than those conducted on multiple species. Although phenotypic patterns have been documented, a major challenge that remains is to resolve the extent to which this variation is due to changes in gene frequencies across populations, to phenotypic plasticity, or epigenetic variation (see below); these mechanisms not only remain unresolved to the extent of invasion, but also in relation to other patterns, such as the latitudinal cline in body size first identified by Johnston and Selander (1964).

Though studies over many decades have focused on invasive populations of the house sparrow to answer evolutionary questions, only relatively recently has attention focused on characterizing the nature of invasions themselves. One of the most comprehensively understood house sparrow introductions and range expansions is that of the Kenyan house sparrow. House sparrows were introduced to Mombasa, on the coast of Kenya in the 1950s (Anderson, 2006) and subsequently expanded their range in a north-westerly direction colonizing most major cities in Kenya by 2007 (Martin et al., 2014); this introduction was most likely a secondary introduction from an introduced population from South Africa (Anderson, 2006). As expected, due to the recency of the introduction, Kenyan house sparrows have reduced genetic diversity compared to other house sparrow populations around the world (both native and those introduced into North America) (Schrey et al., 2011b). Despite an overall reduction in genetic diversity, throughout Kenya, house sparrows display dramatic phenotypic patterns consistent with time since colonization: house sparrows at the range edge are more exploratory (Liebl and Martin, 2012) and more likely to approach and eat novel food (Liebl and Martin, 2014) than individuals from more established areas. Further, at the range edge in Kenya, where stressors are likely less predictable, individuals release more corticosterone in response to stressors (Liebl and Martin, 2012) and have a lower ratio of mineralocorticoid to glucocorticoid receptors (Liebl and Martin, 2013); combined, this likely allows individuals at the range edge to better cope with stressors, whereas individuals in areas of longer establishment mitigate the negative effects of elevated stress hormones. Immunologically, house sparrows at the range edge expressed more Toll-like receptors responsible for identification of parasites, potentially allowing better surveillance of novel parasites (Martin et al., 2014). Although...
individual genetic differences associated with particular
traits were not studied in these studies, some data indi-
cates that epigenetic changes (Liebl et al., 2013a) and
phenotypic flexibility (Liebl and Martin, 2014, Martin
and Liebl, 2014) are at least partially responsible for the
patterns observed throughout this range expansion.

Below, we argue that house sparrows make a particu-
larly adept model species to study invasion and range
expansion of a vertebrate species. Its broad distribution
and relatively well-documented invasion history throu-
ghtout the world alone make house sparrows a valuable
resource for invasion biologists. However, researchers
using house sparrows as a model of invasive species also
benefit from the fact that so much phenotypic and mo-
lecular background and resources already exist. Here, we
focus on and summarize the molecular studies conducted
on house sparrows throughout the world, beginning with
allozymes through to recent work on epigenetic patterns.
We discuss how particular patterns of dispersal through
an introduction event and/or range expansion of this
species have contributed to the genetic patterns observed.
Finally, we discuss the future in invasion genetics and
how using house sparrows as a model for invasion ge-
netics will be particularly enlightening for the field of
invasion biology as well as adaptation generally (Fig. 2).

2 Insights from the House Sparrow

Although not always in the context of invasion or
range expansion, house sparrows have been the focus of
molecular research for many years (Table 1). The investi-
gation of genetic changes across global populations
has and will continue to use cutting edge molecular
tools. For example, the house sparrow was the focus of
the first study of extra-pair parentage using DNA fin-
gerprinting (Burke and Bruford, 1987), some of the ear-
liest studies using microsatellite markers in molecular
ecology (Neumann and Wetton, 1996; Griffith et al.,
1999b), and the more recent use of SNP chip technolo-
gy (Hagen et al., 2013). Most commonly, genetics have
been used to investigate population structure (Table 1)
or behavioral attributes of populations (e.g. extra-pair
paternity (Table 1)). Although most of this work was not
conducted from a perspective of invasion biology or
range expansion, it provides a foundation of knowledge
with validated molecular tools for future work in these
areas.

2.1 Allozymes and microsatellites

Early work using allozymes unfortunately detected
very little variation among populations (Klitz, 1973;
Cole and Parkin, 1981; Parkin and Cole, 1984; Vä-
sänen and Lehväsliho, 1984). These markers, however,
did detect differences among house sparrows across vast
geographic distances: England, Western Europe, Aus-
tralia, and New Zealand were determined to be differen-
tiated (Parkin and Cole, 1985). As would be expected in
introduced populations, house sparrows from Australia
and New Zealand had fewer rare alleles and lower le-
vels of heterozygosity (New Zealand) or higher levels
of genetic differentiation (Australia). This allozyme-
based work set the table for future studies by identifying
a genetic pattern consistent with introduction history
and showing that genetic markers could identify impor-
tant ecological characteristics of these populations.
Population genetic work in house sparrows, however,
expanded significantly with the development of micro-
satellite markers (Neumann and Wetton, 1996; Griffith
et al., 1999b; Richardson et al., 2000; Dawson et al.,
2006; Griffith et al., 2007; Dawson et al., 2010; Dawson

Fig. 2  House sparrows, native to Europe and Asia, have
been introduced all over the world
Among populations, considerable phenotypic variation exists. This
variation is a direct result of many environmental factors influencing
different molecular factors, including bottlenecks, founder effects,
genetic isolation, epigenetic diversity, and phenotypic plasticity and
flexibility.
Table 1  Genetic studies using house sparrows as a model species. Studies are divided by type of study, genetic tool used, the location of the populations studied (I: introduced, N: native), and sample size of each study

| Type of study               | Type of genetic tool used | Population Location | Sample size | Citation                        |
|-----------------------------|---------------------------|---------------------|-------------|---------------------------------|
| Effective population size   | Microsatellites           | Europe (N)          | 6 island populations | Engen et al., 2007. |
| Effective population size   | Microsatellites           | Europe (N)          | 15 populations   | Baalsrud et al., 2014.        |
| Genetic diversity           | Microsatellites           | North America (I)   | 42           | Wetzel et al., 2012.          |
| Genetic diversity           | Allozymes                 | Australia (I)       | 174          | Mannwell and Baker, 1975.     |
| Genetic diversity           | Allozymes                 | Europe (N)          | 133          | Cole and Parkin, 1981.        |
| Genetic diversity           | Allozymes                 | Europe (N)          | 145          | Väisänen and Lehrvänlihio, 1984. |
| Genetic diversity           | Allozymes                 | North America (I)   | 186          | Bates and Zink, 1992.        |
| Genetic diversity           | Microsatellites           | Europe (N)          | 169          | Jensen et al., 2007.        |
| Genetic diversity           | Microsatellites           | Europe (N)          | 355          | Billings et al., 2012.       |
| Marker discovery            | Microsatellites           | Europe (N)          | 40           | Neumann and Wetton, 1996.    |
| Marker discovery            | Microsatellites           | Europe (N)          | 37 - 54      | Griffith et al., 2007.       |
| Marker discovery            | Microsatellites           | Europe (N)          | 64 from 2 populations | Dawson et al., 2012. |
| Parentage                   | DNA-fingerprinting        | North America (I)   | 136          | Whitekiller et al., 2000.    |
| Parentage                   | Microsatellites           | Europe (N)          | 1 population (3 years) | Ockendon et al., 2009. |
| Parentage                   | Microsatellites           | Europe (N); North America (I) | 133 broods from 92 pairs, from 2 populations | Griffith et al., 1999b. |
| Parentage                   | Microsatellites           | North America (I)   | 164          | Stewart et al., 2006.        |
| Parentage                   | Microsatellites           | Europe (N)          | 188 from 5 populations | Liker et al., 2009.         |
| Parentage                   | Microsatellites           | Europe (N)          | 600 from 26 populations | Vangestel et al., 2011b. |
| Parentage                   | Allozymes                 | Europe (N)          | 61           | Töth et al., 2009.          |
| Parentage                   | DNA fingerprinting        | Europe (N)          | 183 broods   | Wetton and Parkin, 1991.     |
| Parentage                   | DNA fingerprinting; allozymes | Europe (N)          | 144 broods   | Wetton et al., 1995.        |
| Parentage                   | Single locus DNA profiling| Europe (N)          | 109          | Cordero et al., 1999.       |
| Parentage                   | Minisatellites; DNA fingerprinting | North America (I) | 528          | Edly-Wright et al., 2007.  |
| Permanent genetic resources | Microsatellites           | Europe (N)          | 621 from 4 populations | Fleischer and Johnston, 1982. |
| Phenotypic association      | Allozymes                 | North America (I)   | 600 from 26 populations | Vangestel et al., 2011a.  |
| Phenotypic association      | Microsatellites           | Europe (N)          | 238 from 10 populations | Kekkonen et al., 2012. |
| Phenotypic association      | Microsatellites           | South America (I)   | 150 from 10 populations | Lima et al., 2012.          |
| Phenotypic association      | Microsatellites           | Europe (N)          | 705 from 12 populations | Brommer et al., 2014.       |
| Population structure        | Allozymes                 | Europe (N)          | 14 populations | Parkin and Cole, 1984.      |
| Population structure        | Allozymes                 | Europe (N)          | 11 populations | Bjordal et al., 1986.       |
| Population structure        | Allozymes; Europe (N); Australia (I); New Zealand (I) | 30 - 160 per population from 21 populations | Parkin and Cole, 1985. |
| Population structure        | Allozymes                 | North America (I)   | 303 from 10 populations (serum) and 122 from 5 populations (tissue) | Klitz, 1973. |
| Population structure        | Allozymes                 | North America (I)   | 447 from 5 populations | Fleischer et al., 1983. |
| Population structure        | Microsatellites           | Africa (I)          | 233 from 10 populations | Schrey et al., 2014. |
| Population structure        | Microsatellites           | Europe (N)          | 4 populations | Hole et al, 2002b. |
| Population structure        | Microsatellites           | Europe (N)          | 472 from 14 populations | Kekkonen et al., 2011b.  |
| Population structure        | Microsatellites           | Europe (N)          | 600 from 26 populations | Vangestel et al., 2012.  |
| Population structure        | Microsatellites           | Europe (N)          | 636 from 14 populations | Jensen et al., 2013. |
| Population structure        | Microsatellites           | North America (I); Europe (N); Africa (I) | 316 from 16 populations | Schrey et al., 2011b. |
| Population structure        | Microsatellites; MHC      | Europe (N)          | 20 - 85 per population from 13 populations | Loiseau et al., 2009. |
| Population structure        | Mitochondrial DNA; 3 nuclear loci | Worldwide | 181 | Sætre et al., 2012. |
| Population structure        | SNP chip                  | Europe (N)          | >160 adults   | Hagen et al., 2003.          |
| Population structure        | Microsatellites           | Europe (N)          | 986 from 2 populations | Kekkonen et al., 2011. |

*Note: Each cell represents the specific genetic tool used for the study, the location of the populations studied, and the sample size. The population structure and phenotypic association studies are highlighted. The citation column lists the references for each study.*
et al., 2012), which have been used to estimate population metrics in house sparrows throughout the world (e.g. heterozygosity, inbreeding coefficient, effective population size, relatedness, and estimates of genetic differentiation (see citations within Table 1)). These markers have proven much more powerful than allozymes at resolving the often subtle patterns of genetic differentiation found among introduced populations of house sparrow.

2.2 Patterns of genetic differentiation

Many ecological factors, some related to invasion and range expansion, affect the genetic differentiation of house sparrows. Native populations show genetic differentiation, even at rather short geographic distances, whereas introduced populations often have less differentiation. In house sparrows, several factors may contribute to the observed genetic differentiation; these include a sedentary lifestyle with low dispersal (Hole et al., 2002), isolation by distance (Loiseau et al., 2009), barriers to migration (Hagen et al., 2013; Jensen et al., 2013), habitat fragmentation (Bjordal et al., 1986; Hole et al., 2002), human commensalism (Bjordal et al., 1986), population declines (Hole et al.; Kekkonen 2011), bottleneck founder effect and genetic drift associated with island colonization (Hagen et al., 2013; Jensen et al., 2013), and biased dispersal over developed habitats (Vangestel et al., 2011a; Vangestel et al., 2012a). In fact, a decline in population size, along with habitat fragmentation and the sedentary nature of house sparrows likely played an important role in generating genetic differentiation among populations within 24 km on farmland in England (Hole et al., 2002) and a striking increase in genetic differentiation was detected after 40 years of population decline among Finnish house sparrow populations (Kekkonen et al., 2011a; Kekkonen et al., 2011b). In addition, introduced populations face several other unique circumstances leading to genetic differentiation including multiple sources of introduced individuals, human-mediated long-distance dispersal, dispersal contingent on introduction history, and multiple patterns of dispersal during range-expansion following the initial introduction (Parkin and Cole, 1985; Schrey et al., 2011a; Lima et al., 2012; Schrey et al., 2014).

Often, there is less genetic differentiation among introduced populations following a range expansion than among native populations over a similar geographic scale. Introduced house sparrows in North America have low genetic variation at allozyme loci (Klitz, 1973) with no genetic structure within a 12 km² area in Kansas (Fleischer, 1983). Further, introduced North American (Schrey et al., 2011) and Brazilian populations (Lima et al., 2011) had lower levels of genetic differentiation compared to native European populations. It is important to note, however, that the introduction of house sparrows into novel areas can generate genetic differentiation via founder effects and genetic drift; house sparrows in Australia and New Zealand (Parkin and Cole, 1985) and Kenya (Schrey et al., 2011), for example, showed genetic differentiation from native populations following introduction.

The studies of recent introductions into Brazil (Lima et al., 2011) and Kenya (Schrey et al., 2011, 2014) suggest genetic characteristics of house sparrow populations change immediately following an introduction event (Schrey et al., 2011; Lima et al., 2012). A global comparison including house sparrow populations from North America, Europe, and Kenya showed that the most recently introduced population (Kenya), had the lowest genetic diversity, whereas the North American populations had seemingly recovered genetically in the 150 years since introduction. In a second, similar study, Brazilian house sparrows (long-established introduced population) had slightly reduced levels of genetic diversity compared to European house sparrows (Lima et al., 2012).

2.3 Patterns of range expansion following an introduction

Interestingly, the genetic characteristics of house sparrow populations also change as the population undergoes range expansion (Lima et al., 2012; Schrey et al., 2014). Depending on the mechanism of range expansion, from serial, stepwise expansions with multiple bottlenecks to multiple introduction events with long distance dispersal, genetic diversity and differentiation can either increase or decrease throughout an expansion (Kolbe et al., 2004; Schrey et al., 2014). Lima et al. (2012) showed little genetic differentiation among Brazilian house sparrows, suggesting expansion occurred from a single, large source population (Lima et al., 2012); at the time of the study, however, house sparrows had been established in Brazil for over 100 years. A second study, addressing population-level genetic changes in Kenyan house sparrows, was conducted less than 60 years from the initial introduction in a population still expanding its range (Schrey et al., 2014). The study found evidence of genetic admixture (local populations including individuals with diverse genetic backgrounds) increasing with range expansion along routes commonly traveled by humans, yet evidence of contemporaneous stepwise expansion in less traveled areas. The pattern of
genetic differentiation observed was consistent with two
types of range expansion. One occurred with multiple
(likely human-mediated) movements of birds between
an introduction source and cities along more traveled
routes. The other occurred via step-wise expansion with
little to no genetic admixture along less well-traveled
human routes.

2.4 Patterns of dispersal

Dispersal is a critical process that influences genetic
structure and likely has a tremendous impact on inva-
siveness. Generally, individual house sparrows do not
disperse far (Fleischer et al., 1984; Altwegg et al., 2000;
Hole et al., 2002b; Skjelseth et al., 2007). Dispersal is
most common among juveniles (Fleischer et al., 1984;
Altwegg et al., 2000), although it is not related to body
mass, condition, clutch size, or hatch date (Altwegg et
al., 2000) suggesting local dispersal may be affected
more by breeding productivity and population size than
nest or genetic effects. Globally, the distribution of
house sparrows tends to be highly fragmented in a
manner consistent with human habitation (Anderson,
2006), and human activity has been implicated in its
current distribution (Anderson, 2006; Schrey et al.,
2014). In particular, the range expansion of introduced
house sparrows in North America and Kenya was highly
influenced by human-mediated long distance dispersal
(Robbins, 1973; Schrey et al., 2011a; Schrey et al.,
2014), which has likely had a profound effect on the
genetic patterns observed in these areas.

Studies have detected barriers to dispersal as having
profound consequences on genetic structure. Within
Finland, for instance, dispersal has maintained connec-
tivity between mainland populations but the sea be-
tween Finland and Sweden, which acts as a dispersal
barrier has led to population differentiation (Kekkonen
et al., 2011). This result is analogous to allozyme results
in Norway, where strong differentiation occurred be-
tween Norway and mainland Europe (Bjordal et al.,
1986). Additionally, island and mainland house sparrow
populations show different patterns of genetic differen-
tiation: island populations tend to have greater differen-
tiation due to bottlenecks, founder effects, and genetic
drift (Hagen et al., 2013; Jensen et al., 2013), but main-
land populations show stronger isolation by distance
(Jensen et al., 2013). Finally, how landscapes are de-
veloped (i.e. degree of urbanization) contributes to dis-
persal patterns, even at small scales (Vangestel et al.,
2012). Urban populations tend to have more fine-scale
structure than rural ones, possibly due to reduced dis-
persal or by attracting greater numbers of migrants from
surrounding suburban and rural areas (Vangestel et al.,
2012). The evolutionary ecology of the urban environ-
ment is an area of growing importance, and not surpris-
singly, being a highly urbanized species, the house
sparrow has been the focus of a number of important
studies in this area. The gradient between the urban and
rural landscape represents a number of ecological chal-
lenge for a species, and the study of behavioral (Bóko-
ny et al., 2009; 2012), physiological (Gavett and Wake-
ley, 1986), and morphological variation (Liker et al.,
2008) across this gradient is highly informative about
adaptiveness, resilience, and invasiveness broadly.

2.5 Beyond microsatellites

One key factor that many of the above described stu-
dies do not address is that microsatellite diversity usually
measures changes in neutral markers rather than diver-
sity of functional genes. To truly understand how inva-
sions are genetically influenced, we will need to transi-
tion from the use of neutral markers to determining how
variation of particular genes influence introduction and
range expansion success. In fact, one study comparing
neutral marker diversity using microsatellites with ma-
jor histocompatibility complex (MHC) divergence show-
ed that MHC differentiation was stronger than that of
the neutral markers (Loiseau et al., 2009); MHC are
some of the most polymorphic loci in vertebrates and
code for a major part of the immune system. Using var-
iation in the functional gene (i.e. MHC), the authors
also reported a stronger association of genetic diversity
with geographic distance than with microsatellites
(Loiseau et al., 2009).

Recent advances in molecular tools (e.g. next genera-
tion sequencing) undoubtedly will provide major con-
tributions to genetic studies of house sparrows. A recent
study (Hagen et al., 2013), which took advantage of the
house finch genome to support SNP-based technology,
demonstrates the promise of this technique, which is
even more powerful than microsatellites. Future studies
will likely take advantage of transcriptomics techniques
(i.e. RNAseq) to assay global pattern of gene expression.
The increased statistical power, the potential to screen
expression data, and the availability of large scale DNA
sequence variation data will allow a deeper investiga-
tion of the genetic factors underlying the well-docu-
mented ability of the house sparrow to invade new en-
vironments. Additionally, these techniques will likely
uncover fundamentally important factors for evoluto-

dary change, which may be applied to other species.

2.6 A role for epigenetic mechanisms

Populations can undergo phenotypic adjustment by
other means than genetic-based evolution to best match variable environments (Pigliucci and Muller, 2010). Phenotypic changes that occur within a population as it transitions from new to colonized have been observed in just a few generations (Duckworth and Badyaev, 2007). This rate of change is unlikely to be explained by genetic evolution alone, especially in introduced populations with reduced genetic diversity. Thus, epigenetic mechanisms or phenotypic plasticity may contribute to changes observed during an introduction or range expansion. Invaders at the vanguard of an expanding wave are exposed to rapidly fluctuating and novel environments; within-individual mechanisms of rapidly adjusting phenotypes to the ensuing environments would be particularly useful in these areas. It is even possible that once phenotypes are determined epigenetically or through plasticity, that canalization of those phenotypes would occur through traditional genetic evolution (Waddington, 1942).

Epigenetic control of phenotypes is one mechanism by which phenotypes are able to adapt rapidly to novel environments. Epigenetic effects induce a change in gene expression without an underlying change in the genomic sequence (Richards, 2006). The most commonly studied epigenetic effect is DNA methylation, which has been shown to generate ecologically important phenotypes (Bossdorf et al., 2010). Epigenetic changes typically increase or decrease expression of particular genes, thus influencing phenotypes. Epigenetically mediated phenotypes are both heritable (Slatkin, 2009) and influenced de novo by environments. Thus, epigenetic variation may be particularly impactful during an introduction event, particularly given the timescale over which introductions and range expansions occur. In other words, as epigenetic marks are influenced by novel environments, they provide a mechanism for modifying, and increasing the expression of particular phenotypes, over and above the underlying genes responsible for the inheritance of that phenotype. Additionally, as the epigenetics marks responsible for expressing (or not) a particular gene can also be heritable, they can themselves be quickly fixed (Waddington, 1942). Indeed, epigenetic variation has been shown to be important in multiple invasions (Richards et al., 2012) including that of house sparrows (Schrey et al., 2012, Liebl et al., 2013a). Compared to a long-established population of introduced house sparrows (Florida, North America), a recently introduced population (Kenya) had more frequent methylation throughout the genome (Schrey et al., 2012). Further, within Kenya, groups with the lowest genetic diversity had the highest epigenetic diversity (Liebl et al., 2013a). This suggests that epigenetic diversity may compensate for the reduced levels of genetic diversity that are often experienced following bottlenecks and founder effects from introduction events.

Although not greatly pursued in the ecology literature, we recommend that future work focus on the importance of epigenetic change for individual fitness and adaptive change during invasions and range expansions. Again, given the distribution of house sparrows and their genetic similarity to passerine species with known genomes, they make ideal candidates to study epigenetic differentiation as it relates to vertebrate invasion. Additional studies addressing broad methylation patterns should be conducted to confirm the patterns in Kenya (that groups with the lowest genetic diversity were those with highest epigenetic diversity; Liebl et al., 2013) are wide-spread among introduced house sparrow populations. In addition, we suggest researchers address methylation and other epigenetic regulatory mechanisms of specific genes. This can be conducted via several techniques including sequencing of genomic material under epigenetic control. This effort would provide crucial information as to how epigenetic mechanisms regulate invasion in house sparrows (and other species). Potential target genes which have been identified as under epigenetic control in other species include those influence stress regulation (Weaver et al., 2004), beak shape (Skinner et al., 2014), and immune regulation (Skinner et al., 2014).

3 Future Directions

To best understand invasion genetics, we believe a global collaboration by a consortium of researchers should join forces to ask questions in a coherent manner at an international level; further, we believe the house sparrow is one of the better models for such work. The house sparrow provides excellent opportunities for the field of invasive species genetics because of its distribution, abundance, and considerable background of knowledge. These characteristics prevent problems caused by low replication or sample size, at both the individual and population level. Borer et al. (2014) present the success that is achievable by an international consortium studying a single research question. Studies of globally distributed house sparrows would provide insight not only into the genetics of the invasion, but also to other major questions of global significance such as those related to biodiversity, conservation, and organismal adaptability to changing environments.
Conservation biology has focused on the study of rare species, but relatively few on abundant species expanding their range. However, focusing on these common species and what traits confer resilience will lend an important perspective to our effort of preserving biodiversity (Gaston, 2010). Unarguably, house sparrows are common throughout its range, although, ironically, may be at a lower density in parts of its native range (e.g. UK) than introduced range (Hole et al., 2002). There are a host of questions regarding the characteristics that have conferred such success on the house sparrow as an invasive species. To what extent can this success be attributed to plasticity in reproduction, diet, sociality, cognitive skills, immunity, or is there something about the genetic architecture of the species that has enabled it to respond so rapidly to such a wide range of environments?

It may be argued that house sparrow invasion success is largely dependent on the fact they are predominantly found in human modified areas, which is becoming increasingly distributed across the globe. However, it is also clear that the species is able to survive (and thrive) in variable climates, including those very different from its native range. This provides an ideal opportunity to use house sparrows across the world to understand adaptation to different and changing climates. For example, house sparrows introduced from England have become established along a clinal gradient in Australia from temperate Tasmania to monsoonal tropics of Queensland through the arid interior. Addressing variation in behavioral, physiological, morphological, and molecular traits across such clines would deliver significant insight into how climatic variation drives selection on avian biodiversity, as well as providing a detailed understanding about the mechanisms through which birds can adapt to climatic variation at an almost global level. Whereas these ideas are not specifically genetic, the addition of modern molecular techniques in conjunction with such studies would provide great additional insight such as to what extent genetic or epigenetic variation determine the responses of individuals and/or populations to different environments.

In conclusion, due to its widespread distribution and up-and-coming molecular tools, house sparrows make an ideal model in which researchers can study invasion and range expansion. The factors deemed important in the successful invasion potential of house sparrows should prove to be important for management decisions concerning invasive species as well as for conservation of threatened species undergoing range expansions and shifts due to environmental changes. Because house sparrows have been independently introduced all over the world, researchers can identify the molecular correlates of invasion and range expansion without the problems experienced in singly introduced species. Given the global distribution pattern of house sparrows, they also present the opportunity for researchers around the world to collaborate on a similarly focused goal, of identifying the genetic mechanisms of range expansion. Finally, those characters identified as major contributors to invasion and range expansion success in the house sparrow likely are not unique to the house sparrow. However, the identification of such characters in such a well-developed invasion trajectory allows research to identify key traits important to invasions that can be pursued in other, invasive species.

Acknowledgements We would like to thank Lee Ann Rollins for organizing the special column and inviting this submission. ALL was supported by National Environment Research Council (NE/K005766) while this manuscript was being written. AWS was supported by the College of Science and Technology and Department of Biology at ASU. SCA and ES were supported by Macquarie University Research Excellence Scholarships. SCG was supported by an Australian Research Council Future Fellowship (FT13D101253).

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