Uptake, translocation and biotransformation kinetics of BDE-47, 6-OH-BDE-47 and 6-MeO-BDE-47 in maize (Zea mays L.)

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Abstract
This study presents a detailed kinetic investigation on the uptake, acropetal translocation and transformation of BDE-47, 6-OH-BDE-47 and 6-MeO-BDE-47 in maize (Zea mays L.) by hydroponic exposure. Root uptake followed the order: BDE-47 > 6-MeO-BDE-47 > 6-OH-BDE-47, while 6-OH-BDE-47 was the most prone to acropetal translocation. Debromination rates of BDE-47 were 1.31 and 1.46 times greater than the hydroxylation and methoxylation rates, respectively. Transformation from BDE-47 to lower brominated OH/MeO-PBDEs occurred mainly through debromination first followed by hydroxylation or methoxylation. There was no transformation from 6-OH-BDE-47 or 6-MeO-BDE-47 to PBDEs. Methylolation rate of 6-OH-BDE-47 was twice as high as that of 6-MeO-BDE-47 hydroxylation, indicating methylation of 6-OH-BDE-47 was easier and more rapid than hydroxylation of 6-MeO-BDE-47. Debromination and isomerization were potential metabolic pathways for 6-OH-BDE-47 and 6-MeO-BDE-47 in maize. This study provides important information for better understanding the mechanism on plant uptake and transformation of PBDEs.

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1. Introduction
Polybrominated diphenyl ethers (PBDEs) have been widely used as additive flame retardants in a variety of household and industrial products for decades (de Wit, 2002). Based on the knowledge of their adverse effects on the environment and human health, some PBDEs have been included on the Stockholm Convention list of priority persistent organic pollutants, and some technical products have been banned or voluntarily withdrawn from use in some regions of the world (California State Assembly, 2003; Cox and Efthymiou, 2003). However, owing to their high volume production, long-term use, and lipophilic and persistent properties, PBDEs are still frequently detected in various environmental media (Kim et al., 2014; Newton et al., 2015) and biotic samples (Shang et al., 2013; Wang et al., 2014; Zhu et al., 2009) in the environment. Therefore, there is considerable concern about their behaviors and fates in the environment.

In addition, PBDEs have a tendency to break down into lower brominated congeners in the environment (Gandhi et al., 2011; He et al., 2006; Tokarz et al., 2008). Hydroxylated PBDEs (OH-PBDEs) and methoxylated PBDEs (MeO-PBDEs) have also been identified in various environmental media (Ueno et al., 2008; Wang et al., 2014) and biological samples (Kelly et al., 2008; Teuten et al., 2005). Their origin is far from clear, and increasing evidence has shown the existence of transformation of PBDEs to OH-PBDEs and MeO-PBDEs (Stapleton et al., 2009; Sun et al., 2013a; Wang et al., 2012). Debrominated PBDEs, OH-PBDEs and MeO-PBDEs have greater potencies for some adverse effects such as embryo developmental toxicity (Boxtel et al., 2008), genotoxicity (Ji et al., 2011) and endocrine disrupting effects (He et al., 2008; Meerts et al., 2001; Wiseman et al., 2011) than their precursor PBDEs, and may bring additional adverse influences to bear on the environment and human health; therefore their formations and behaviors in the environment have received increasing attention.

Soils represent a major sink for organic contaminants in the environment, and PBDEs have been frequently detected in soils (Tang et al., 2014; Wang et al., 2014). For example, a total concentration of PBDEs ranging between 65 and 6080 ng kg⁻¹ dry weight was observed in UK and Norwegian soils in a survey of European soils (Hassanin et al., 2004). In our previous study, we identified a total of 41 PBDE congeners, with the total concentrations ranging
from 13.9 to 13, 251.2 ng g⁻¹ dry weight, together with twelve OH-PBDEs and MeO-PBDEs in the soils collected in Qingyuan county in Guangdong Province (Wang et al., 2014). Concentrations of PBDEs in plants were even reported as high as 70–5900 ng g⁻¹ dry weight by a field investigation in Liaozhou, China (Jin et al., 2008). The uptake, translocation and accumulation of PBDEs have been demonstrated by previous studies (Huang et al., 2011; Wang et al., 2014). Root concentration factors (RCFs) of PBDEs were found 0.02–2.18 and 0.01–1.2 for plants uptake from e-waste soils in a pot experiment (Huang et al., 2011) and field experiment (Wang et al., 2014), respectively. Furthermore, metabolic transformation of organic contaminants in plants is an important biotransformation process in terrestrial ecosystems (Sandermann, 1994; Yu et al., 2013). Despite their high importance, studies on the behavior of PBDEs in terrestrial plants are very limited, particularly regarding the biotransformation of PBDEs, OH-PBDEs and MeO-PBDEs in plants. In our previous work, transformation of PBDEs to lower brominated PBDEs, OH-PBDEs and MeO-PBDEs in plants was observed (Huang et al., 2010; Wang et al., 2012). Sun et al. (2014) recently reported the reciprocal transformation between OH-PBDEs and the corresponding MeO-PBDEs in young whole pumpkin plants by a hydroponic experiment. Unfortunately, this study only focused on the inter-conversions between MeO-PBDEs and corresponding OH-PBDEs without considering the possibility of further transformation. More importantly, there is a paucity of research on the transformation kinetics of PBDEs, OH-PBDEs and MeO-PBDEs, which is necessary to gain a thorough insight into their metabolic pathways.

The aim of this study was to investigate the uptake, translocation and biotransformation of PBDEs, OH-PBDEs and MeO-PBDEs by maize. PBDEs always exist as a mixture in field soils, which brings difficulties to distinguish the debranified products from the same PBDE congeners, originally existed in the soils. Pot experiment with soil will introduce the influences of soil microorganisms on biotransformation and bring difficulties for elucidating biotransformation of PBDEs by plants. Therefore, a hydroponic experiment was conducted although it cannot entirely exclude the influences from root-associated microbes. BDE-47 is frequently detected in PBDE-contaminated soils and accumulated in plants at relatively high concentrations (Kim et al., 2014; Sun et al., 2013b). So, BDE-47 and its hydroxylated and methoxylated analogs (6-OH-BDE-47 and 6-MeO-BDE-47) were selected as the target compounds, and their kinetic uptake by roots and acropetal translocation in maize were investigated. Biotransformation products within maize after exposure to BDE-47, 6-OH-BDE-47 and 6-MeO-BDE-47 at different time intervals were identified. Based on the kinetic profiles of metabolic congeners and their concentrations, the metabolic pathways of BDE-47, 6-OH-BDE-47 and 6-MeO-BDE-47 in maize are discussed.

2. Material and methods

2.1. Chemicals

Chemical standards of BDE-47, 6-MeO-BDE-47 and 6-OH-BDE-47 (purity > 98.0%, New Haven, CT, USA) were used as exposure compounds (Table S1). ¹³C-PCB-141 and ¹³C-6-OH-BDE-47 were used as surrogate standards purchased from Cambridge Isotope Laboratory (Andover, MA, USA). A mixed standard solution of PBDEs containing 24 PBDEs through mono-to tetra-BDEs, standards of nine OH-PBDEs and standards of nine MeO-PBDEs from Accusstandard (New Haven, CT, USA) were used in the identification and quantification of the metabolites. All these compounds were analyzed in the BDE-47, 6-MeO-BDE-47 and 6-OH-BDE-47 solutions to check their purity, and the presence of diverse impurities in solutions (Table S2) did not affect the exposure experiments as well as discussion on metabolic reactions. Distilled water was used in all the experiments. Solvents, including hexanes, acetone, acetonitrile, dichloromethane (DCM) and methyl tert-butyl ether (MTBE), were of HPLC grade (Thermo fisher, MA, USA). All the other chemicals used were of analytical grade (Sinopharm Chemical Reagent Co., Ltd, BJ, China).

2.2. Exposure experiment

The exposure solutions of BDE-47, 6-OH-BDE-47 and 6-MeO-BDE-47 were obtained by first dissolving the standard solution in acetone, and then gradually diluting with sterile half-strength Hoagland nutrient solution in bottles wrapped with aluminum-foil paper. The nominal concentrations of BDE-47, 6-OH-BDE-47 and 6-MeO-BDE-47 in the exposure solutions were set at 20 nmol L⁻¹. The volume of acetone in the test solutions was less than 1% (v/v). The pH was adjusted to 6.5. Each solution was prepared immediately prior to exposure and mixed thoroughly to ensure complete dissolution. The concentrations of BDE-47, 6-OH-BDE-47 and 6-MeO-BDE-47 in the solutions were then determined at 18.4 ± 1.1, 17.9 ± 1.7 and 18.9 ± 0.9 nmol L⁻¹ before exposure, respectively.

Maize (Zea mays L.) seeds were obtained from the Chinese Academy of Agricultural Sciences, Beijing, China. Prior to germination, seeds of similar size were selected and surface-sterilized in 3% (v/v) H₂O₂ for 30 min, followed by thoroughly washing with distilled water, and subsequently germinated on moist filter paper. After 4 days, uniformly (about 5 cm in height) germinated seedlings were transferred to glass containers containing half-strength Hoagland nutrient solution and ready for exposure after 7 d of cultivation.

The whole seedlings were transferred to sterile glass-stoppered flasks for the exposure, which were wrapped with black paper to eliminate photolysis of the compounds. Then 150 mL of the test solutions of BDE-47, 6-OH-BDE-47 and 6-MeO-BDE-47 were added individually. Treatments with maize seedlings in 150 mL of sterile half-strength Hoagland nutrient solution were set as untreated controls to monitor possible cross contamination, and treatments with 150 mL test solution but no plants were set as unplanted controls to monitor any possible loss. Pots were kept in a controlled environment growth chamber at a light intensity of 250 μmol m⁻² s⁻¹ provided by supplementary illumination with a photoperiod of 14 h each day, at a 25/20 °C day/night temperature regime and a relative humidity of 80%. Approximately 20 mL d⁻¹ of sterile half-strength Hoagland nutrient solution saturated with oxygen was injected into each container to compensate for transpiration losses. Maize seedlings were harvested at intervals of 3, 6, 12, 24, 48, 72 and 96 h. At the end of the exposure, unplanted and untreated controls were sampled. All the samples were prepared in triplicate in separate containers including the ones for different time intervals.

The roots of samples were first thoroughly rinsed with deionized water and blotted with tissue paper, and maize seedlings were separated into roots and shoots for the subsequent analysis. All the harvested plant samples were freeze-dried at –50 °C for 48 h in a lyophilizer (FD-1, Boyikang Instrument Ltd, Beijing, China), weighed and stored at –20 °C for analysis. The subsequent calculations were therefore based on dry weight.

2.3. Sample extraction and analysis

To prevent photolysis of PBDEs, OH-PBDEs and MeO-PBDEs, all the glasses used for sample extraction and cleanup were wrapped with aluminum-foil paper. Before extraction, surrogate standards of
13C-PBDE-141 and 13C-6-OH-BDE-47 (20 ng of each) were added to the samples, and kept for 1 h. Extraction and analysis of PBDEs and MeO-PBDEs in plant samples were based on the methods of López et al. (2009) and Wang et al. (2011b). An Agilent 7890 GC–MS (5975 inert) (Agilent, Palo Alto, CA, USA) and a DB-5MS column (30 m × 0.32 mm i.d., 0.25 μm film thickness, J & W Scientific, Folsom, CA, USA) were used for the analysis of PBDEs and MeO-PBDEs, respectively (Wang et al., 2011b). For the determination of OH-PBDEs, the extraction and cleanup procedures were based on the methods described in our previous research (Wang et al., 2011a) with minor modifications. A UPLC-MS/MS system was used to detect and quantify the OH-PBDEs. Chromatographic separation of OH-PBDEs was performed on an UPLC (Waters ACQUITY UPLC system, Waters, Milford, MA, USA). Reversed-phase chromatography was performed by a Waters ACQUITY UPLC BEH C18 column (2.1 mm × 100 mm, 1.7 μm particle size, Waters, Milford, MA, USA) maintained at 40 °C. Mass spectrometry analysis was carried out using a Waters Xevo TQ MS triple quadrupole mass spectrometer equipped with an ESI source (Waters, Milford, MA, USA). Details of sample extraction and analysis are provided in the supplementary material.

2.4. Quality control and quality assurance

Quality control was done by regular analyses of the procedural blanks, blind duplicate samples, and random injection of solvent blanks and standards. Quality assurance was guaranteed by the addition of surrogate standards and the standards of PBDEs, OH-PBDEs and MeO-PBDEs individually to the blank plant samples. Recoveries were 72.3–97.1% for 13C-PBDE-141 and 77.4–96.5 for 13C-6-OH-BDE-47, and 79.8–110.5%, 70.2–91.3% and 66.2–99.7% for PBDEs, OH-PBDEs and MeO-PBDEs, respectively. The limits of detection (LODs) defined as a signal-to-noise ratio (S/N) of 3 were in the range of 25–52 pg g⁻¹ for PBDEs, 39–83 pg g⁻¹ for MeO-PBDEs and 50–100 pg g⁻¹ for OH-PBDEs in the plant samples (n = 6), respectively. Details of the LODs and matrix spike recoveries for the individual PBDE, MeO-PBDE and OH-PBDE congeners are provided in Table S3 of the supplementary material.

2.5. Modeling

A pseudo-first order kinetic model was applied to assess the metabolism kinetics of BDE-47, 6-OH-BDE-47 and 6-MeO-BDE-47 inside maize, which can be expressed as follows:

\[ q_t = q_m [1 - \exp(-kt)] \]

where \( q_m \) and \( q_t \) (pmol pot⁻¹) are the maximum and the amount of each kind of metabolites generated in maize (the sum of respective debromination, hydroxylation, methoxylation, methylation and isomerization metabolites in maize roots and shoots from one pot) over the exposure time \( t \) (h), respectively, and \( k \) (h⁻¹) is the metabolism rate constant. The data for the amounts of each kind of metabolites generated in maize were plotted in the form of \( q_t \) versus \( t \) for determining first-order kinetics for BDE-47, 6-MeO-BDE-47 and 6-OH-BDE-47, respectively. The fitting parameters were calculated with OriginLab Corporation.

2.6. Data analysis

All data were subjected to statistical analysis using Microsoft Excel 2010 and Origin 8.0 software. Means and standard deviation were calculated for triplicate samples and the results were expressed on a dry weight basis. A T-normality test was conducted for the obtained data before parametric statistics, and the result showed that the data fitted normal distribution. One-way analysis of variance with Duncan’s multiple-comparison test was used to assess differences among samples at \( p < 0.05 \).

3. Results and discussion

3.1. Uptake and translocation of BDE-47, 6-OH-BDE-47 and 6-MeO-BDE-47 by maize

Time-dependent accumulation of BDE-47, 6-OH-BDE-47 and 6-MeO-BDE-47 in maize roots and shoots is displayed in Fig. 1. There was a similar variation tendency for the accumulations of BDE-47 and 6-MeO-BDE-47 in maize, increasing initially until the exposure times of 6 h and 24 h for roots and shoots, respectively, and then decreasing. The concentrations of BDE-47 and 6-MeO-BDE-47 in roots were comparable, while the concentrations of 6-MeO-BDE-47 in shoots were 1.6–5.1 times as high as those of BDE-47. However, distinctly different concentration changes over time were observed for 6-OH-BDE-47 accumulation in maize. The highest concentration of 6-OH-BDE-47 in maize roots was observed at the time of first sampling and then decreased with increasing time; whereas its concentration in shoots increased continuously starting at the exposure time of 24 h. Another difference is that the concentration of 6-OH-BDE-47 in maize roots and shoots was approximately 10–100 fold lower than those of BDE-47 and 6-MeO-BDE-47. In order to compare their uptake abilities, root concentration factors (RCFs) were calculated as the ratio of the concentrations in maize roots to the concentrations in the exposure solutions. The average RCF values were in the following order: BDE-
47 > 6-MeO-BDE-47 > 6-OH-BDE-47, which was consistent with the previous observations in field contamination investigations (Sun et al., 2013b; Wang et al., 2014). A significant positive correlation existed between the values of logRCF and logKow ($R^2 = 0.91$, $p < 0.0001$) (Fig. S1 A), indicating that hydrophobicity determines plant uptake of BDE-47, 6-OH-BDE-47 and 6-MeO-BDE-47 and suggesting that partitioning is the key mechanism for their plant uptake (Chiou et al., 2001; Wang et al., 2011c).

The accumulations of BDE-47, 6-OH-BDE-47 and 6-MeO-BDE-47 in shoots indicated their acropetal translocation in maize. Translocation factors (TF, ng g$^{-1}$ shoot/ng g$^{-1}$ root on dry weight basis) were calculated, and the average TF values followed the order: 6-OH-BDE-47 > 6-MeO-BDE-47 > BDE-47, which had a significantly negative correlation with their logKow values ($R^2 = 0.44$, $p < 0.001$) (Fig. S1 B). There were analog-specific differences in acropetal translocation of BDE-47, 6-OH-BDE-47 and 6-MeO-BDE-47 in maize. The lower logKow value leads to much easier root-to-shoot translocation of 6-OH-BDE-47 than those of BDE-47 and 6-MeO-BDE-47. A reasonable explanation is that the presence of the hydroxyl group in 6-OH-BDE-47 increases its polarity and water solubility and renders it more prone to transfer from roots to shoots than either the corresponding unsubstituted BDE-47 or 6-MeO-BDE-47. Therefore, the translocation of 6-OH-BDE-47 than those of BDE-47 and 6-MeO-BDE-47 isomers was mainly the maize that transformed BDE-47, 6-OH-BDE-47 and 6-MeO-BDE-47 and suggested that partitioning is the key mechanism for their plant uptake (Sun et al., 2013a).

In the BDE-47 exposed maize, besides 6-OH-BDE-47, the lower brominated hydroxylated metabolites, 3′-OH-BDE-28 and 2′-OH-BDE-3, were detected as the dominant species, followed by hydroxyl group or bromine atom rearrangement products of 5-OH-BDE-47 and 2′-OH-BDE-68 (Fig. 2A-b). 6-MeO-BDE-47 was the predominant methoxylated metabolite in maize roots and shoots, followed by 6-MeO-BDE-47 and 3′-MeO-BDE-49 of methoxyl group or bromine atom rearrangement products, and then 3′-MeO-BDE-28 and 2′-MeO-BDE-3 of the lower brominated methoxylated metabolites (Fig. 2A-c). Lower brominated hydroxylated and methoxylated metabolites may be formed through hydroxylation and methoxylation of the PBDE debrominated products or debromination of the hydroxylated and methoxylated products. However, it is still unclear which one is the main transformation pathway.

In order to elucidate the formation pathways of lower brominated hydroxylated and methoxylated metabolites, the metabolic products of 6-OH-BDE-47 and 6-MeO-BDE-47 in maize were determined. Neither BDE-47 nor lower brominated PBDEs was detected in tissues of maize after exposure to 6-OH-BDE-47 or 6-MeO-BDE-47. 6-MeO-BDE-47 was detected for 6-OH-BDE-47 exposure and 6-OH-BDE-47 was found for 6-MeO-BDE-47 exposure (Fig. 2B-a, c-b), suggesting the inter-conversion between 6-MeO-BDE-47 and 6-OH-BDE-47 by maize. Besides their inter-conversion, five methoxylated metabolites (2′-MeO-BDE-68, 5-MeO-BDE-47, 3′-MeO-BDE-28, 3′-MeO-BDE-7 and 2′-MeO-BDE-3) and four hydroxylated metabolites (5-OH-BDE-47, 2′-OH-BDE-68, 3′-OH-BDE-28 and 2′-OH-BDE-3) were detected in the tissues of maize exposed to 6-OH-BDE-47 and 6-MeO-BDE-47, respectively (Fig. 2B-a and c-b). Two lower brominated hydroxylated metabolites, 3′-OH-BDE-28 and 2′-OH-BDE-3, and three lower brominated methoxylated metabolites, 3′-MeO-BDE-28, 3′-MeO-BDE-7 and 2′-MeO-BDE-3, were detected in the tissues of maize exposed to 6-OH-BDE-47 and 6-MeO-BDE-47, respectively (Fig. 2B-b and C-a), suggesting that lower brominated hydroxylated and methoxylated metabolites could be formed through debromination of OH-PBDEs and MeO-PBDEs. Similar results of OH-tetraBDE formation from OH-pentabDE were reported previously (Wan et al., 2009). The isomers of 6-OH-BDE-47 and 6-MeO-BDE-47 resulting from the rearrangement of bromine atoms, hydroxyl or methoxy groups were also detected in maize. For example, 3′-OH-BDE-47, 5-OH-BDE-47 and 2′-OH-BDE-68 were isomers of 6-OH-BDE-47, while 2′-MeO-BDE-68 was an isomer of 6-MeO-BDE-47 (Fig. 2B-b and C-a). Simultaneous detections of 2′-OH-BDE-68, 5-OH-BDE-47, 3′-OH-BDE-47 and 6-OH-BDE-47, 2′-MeO-BDE-68 and 6-MeO-BDE-47 in maize (Qu et al., 2007), fish (Munsch et al., 2010) and maize (Wang et al., 2012) have been reported, supporting our observation on biosomerization of 6-OH-BDE-47 and 6-MeO-BDE-47 in maize.
quantitatively assess the transformation processes of BDE-47, 6-OH-BDE-47 and 6-MeO-BDE-47 in maize. The amounts of different types of metabolic products (debrominated, methoxylated, hydroxylated and isomerized, respectively) generated in maize after exposure at different time intervals were calculated and the results are provided in Fig. S2 in the supplementary material. The data fitted well with the pseudo-first order kinetic model with high $R^2$ values (0.9128–0.9845) and low mean weighted square errors (Table 2). From the fitting results we can see that the values of transformation rate constant ($k$) and the maximum amount of metabolites ($q_m$) generated for BDE-47 debromination were significantly higher than those of its methoxylation and hydroxylation, suggesting that debromination of BDE-47 was easier than its methoxylation and hydroxylation. Moreover, the debromination rate constant of BDE-47 was 5.1 and 2.7 times as high as those of 6-OH-BDE-47 and 6-MeO-BDE-47, respectively, suggesting that debromination of BDE-47 was faster than debromination of 6-OH-BDE-47 and 6-MeO-BDE-47.

The $q_m$ and $k$ values of 6-OH-BDE-47 methylation were 40.7 and 2.8 times as high as those of 6-OH-BDE-47 debromination, respectively, while the $q_m$ and $k$ values of 6-MeO-BDE-47 debromination were 10.3 and 1.4 times as high as those of 6-MeO-BDE-47 hydroxylation. These suggested that methylation and debromination are the main transformation pathways of 6-OH-BDE-47 and 6-MeO-BDE-47, respectively. Though the debromination of hydroxylated and methoxylated products was demonstrated, the higher $q_m$ and $k$ values of BDE-47 debromination compared with those of 6-OH-BDE-47 and 6-MeO-BDE-47 (Table 2) strongly suggested that the formation of lower brominated hydroxylated and methoxylated metabolites from BDE-47 mainly followed the pathway of debromination first and then hydroxylation and methoxylation. The rate of transformation from 6-OH-BDE-47 to MeO-PBDEs was twice as
high as that from 6-MeO-BDE-47 to OH-PBDEs, suggesting that 6-OH-BDE-47 and 6-MeO-BDE-47 were low, coexistence of their isomers in biotic (Lacorte and Ikonomou, 2009; Malmvärn et al., 2008; Qiu et al., 2009) and abiotic (Kelly et al., 2008; Wang et al., 2014; Zhang et al., 2012) media suggested the formation of OH-PBDEs and MeO-PBDEs, through isomerization could not be ignored.

3.4. Metabolic pathways of BDE-47, 6-OH-BDE-47 and 6-MeO-BDE-47 in maize

Table 1

| Compounds | T[h]  | Solution (%) | Accumulation | Metabolism | Recovery (%) |
|-----------|-------|--------------|--------------|------------|--------------|
|           |       | Root (%)     | Shoot (%)    | De bromination (%) | Methoxylation (%) | Hydroxylation (%) | Isomerization (%) |
| BDE-47    | 3     | 14.8 ± 1.1   | 93.7 ± 13.1  | 3.4. Metabolic pathways of BDE-47, 6-OH-BDE-47 and 6-MeO-BDE-47 in the experiment. | 0.007 ± 0.001      | 0.031 ± 0.04     | 0.079 ± 0.03     | 111.8 ± 13.2 |
|           | 6     | 8.50 ± 0.5   | 100.2 ± 9.8  | 4.01 ± 0.7  | N.D.         | 2.32 ± 0.9     | 0.21 ± 0.07     | 0.033 ± 0.08 |
|           | 12    | 2.55 ± 0.1   | 95.2 ± 10.4  | 4.06 ± 0.1  | N.D.         | 2.92 ± 0.08    | 0.71 ± 0.1      | 0.033 ± 0.08 |
|           | 24    | 2.27 ± 0.2   | 81.1 ± 6.3   | 2.02 ± 0.1  | N.D.         | 3.39 ± 0.5     | 1.08 ± 0.2      | 0.054 ± 0.1  |
|           | 48    | 2.11 ± 1.1   | 96.0 ± 9.1   | 4.31 ± 0.3  | N.D.         | 5.76 ± 0.7     | 1.74 ± 0.4      | 0.090 ± 0.07 |
|           | 72    | 1.24 ± 0.4   | 110.0 ± 8.9  | 4.26 ± 0.6  | N.D.         | 1.95 ± 0.1     | 0.078 ± 0.02    | 0.034 ± 0.1  |
|           | 96    | 1.26 ± 0.3   | 98.0 ± 8.7   | 4.92 ± 0.9  | N.D.         | 5.98 ± 0.6     | 1.51 ± 0.2      | 0.079 ± 0.03 |
| 6-OH-BDE-47 | 3     | 69.1 ± 7.1   | 7.91 ± 0.9   | 0.61 ± 0.0  | 0.07 ± 0.002 | 2.17 ± 0.6     | N.D.          | 0.12 ± 0.05 |
|           | 6     | 24.7 ± 3.4   | 5.92 ± 1.1   | 0.61 ± 0.0  | 0.09 ± 0.004 | 4.85 ± 1.0     | N.D.          | 0.22 ± 0.03 |
|           | 12    | 14.2 ± 2.3   | 5.51 ± 0.7   | 0.60 ± 0.1  | 0.14 ± 0.001 | 7.79 ± 0.7     | N.D.          | 0.28 ± 0.03 |
|           | 24    | 11.4 ± 1.9   | 5.89 ± 0.5   | 0.82 ± 0.09 | 0.19 ± 0.03  | 11.5 ± 1.4     | N.D.          | 0.29 ± 0.05 |
|           | 48    | 10.1 ± 1.7   | 3.58 ± 0.2   | 1.05 ± 0.2  | 0.22 ± 0.02  | 12.6 ± 2.3     | N.D.          | 0.32 ± 0.04 |
|           | 72    | 5.87 ± 1.0   | 4.08 ± 0.7   | 2.06 ± 0.2  | 0.49 ± 0.06  | 25.1 ± 3.1     | N.D.          | 0.67 ± 0.1  |
|           | 96    | 6.45 ± 0.9   | 2.41 ± 0.3   | 3.30 ± 0.4  | 0.54 ± 0.04  | 24.6 ± 2.7     | N.D.          | 0.73 ± 0.1  |
| 6-MeO-BDE-47 | 3     | 15.1 ± 4.4   | 99.2 ± 8.8   | 5.17 ± 0.9  | N.D.         | N.D.          | 0.027 ± 0.005 | 0.033 ± 0.001 |
|           | 6     | 11.0 ± 2.3   | 88.2 ± 9.2   | 5.61 ± 1.3  | 0.06 ± 0.001 | 10.4 ± 1.4     | N.D.          | 0.104 ± 0.107 |
|           | 12    | 4.17 ± 0.7   | 98.4 ± 11.7  | 6.52 ± 2.0  | 0.32 ± 0.02  | 11.9 ± 1.4     | N.D.          | 0.104 ± 0.107 |
|           | 24    | 2.53 ± 0.3   | 101.8 ± 5.3  | 8.78 ± 1.5  | 0.53 ± 0.1   | 12.9 ± 1.4     | N.D.          | 0.104 ± 0.107 |
|           | 48    | 6.57 ± 0.5   | 99.5 ± 8.1   | 9.62 ± 2.1  | 0.65 ± 0.3   | 16.7 ± 2.3     | N.D.          | 0.104 ± 0.107 |
|           | 72    | 5.25 ± 0.3   | 93.6 ± 10.6  | 9.03 ± 1.1  | 1.01 ± 0.5   | 20.7 ± 2.3     | N.D.          | 0.104 ± 0.107 |
|           | 96    | 5.79 ± 0.4   | 104.6 ± 7.3  | 9.64 ± 1.6  | 1.17 ± 0.2   | 25.8 ± 2.3     | N.D.          | 0.104 ± 0.107 |

Table 2

| Compounds | Metabolism | qm (pmol pot⁻¹) | k [h⁻¹] | R² | P value | MWSE² |
|-----------|------------|-----------------|---------|----|---------|------|
| BDE-47    | Debromination | 249 ± 19.3      | 0.051 ± 0.010e | 0.9378        | <0.0001 | 0.009 |
|           | Methoxylation | 7.08 ± 7.62     | 0.039 ± 0.012d | 0.9202        | <0.0001 | 0.027 |
|           | Hydroxylation | 3.21 ± 0.21     | 0.035 ± 0.010cd | 0.9542        | <0.0001 | 0.011 |
|           | Methylation  | 4.96 ± 1.23     | 0.010 ± 0.005a | 0.9333        | <0.0001 | 0.013 |
|           | Isomerization | 6.89 ± 2.12     | 0.020 ± 0.007b | 0.9271        | <0.0001 | 0.003 |
| 6-OH-BDE-47 | Debromination | 5.17 ± 6.23     | 0.019 ± 0.008b | 0.9245        | <0.0001 | 0.037 |
|           | Hydroxylation | 5.01 ± 1.05     | 0.014 ± 0.007ab | 0.9479        | <0.0001 | 0.009 |
|           | Isomerization | 2.29 ± 0.59     | 0.010 ± 0.007a | 0.9128        | <0.0001 | 0.044 |

a Metabolism types of each exposure compound.
b Calculated data from the model.
c Different letters represent significant differences between rates of metabolisms for the exposure compounds (P < 0.05; Duncan's test).
d Mean weighted square error, equal to 1/|V-1|Σ[(qexp - qcalc)^²]/qexp, where V is the amount of freedom; v = N - 2 for Eq. (1) (Wen et al., 2011).
Fig. 3. Metabolic pathways of (A) BDE-47, (B) 6-OH-BDE-47 and (C) 6-MeO-BDE-47 in maize.
were also expected to occur in maize. Based on the above analysis, brief description of the metabolic pathways of BDE-47, 6-OH-BDE-47 and 6-MeO-BDE-47 in maize can be drawn as shown in Fig. 3.

4. Conclusions

In summary, BDE-47, 6-OH-BDE-47 and 6-MeO-BDE-47 were accumulated in maize and translocated from roots to shoots over exposure time. Root uptake followed the order: BDE-47 > 6-MeO-BDE-47 > 6-OH-BDE-47, while 6-OH-BDE-47 was more prone to acropetal translocation inside maize than BDE-47 and 6-MeO-BDE-47. Transformations from BDE-47 to OH/MeO-PBDEs, from 6-OH-BDE-47 to MeO-PBDEs and from 6-MeO-BDE-47 to OH-PBDEs took place, but there was no transformation from 6-OH-BDE-47 or 6-MeO-BDE-47 to PBDEs in maize. The dehydration rate constant for BDE-47 was higher than those of its hydroxylation and methoxylation, thus formation of lower brominated OH-PBDE and MeO-PBDE metabolites from BDE-47 occurred mainly via hydroxylation and methoxylation of the debrrominated products of BDE-47. Methylation of 6-OH-BDE-47 occurred more rapidly than hydroxylation of 6-MeO-BDE-47 in maize. Besides methylation and hydroxylation, debrromination and isomerization were proposed as the potential metabolism pathways to form MeO-PBDEs and OH-PBDEs from 6-OH-BDE-47 and 6-MeO-BDE-47 in maize. The results of this study provide valuable information for a better understanding of the phytoaccumulation and phytotransformation of PBDEs, OH-PBDEs and MeO-PBDEs. Further investigation is necessary to clarify such processes in the natural environment. Moreover, the evidences of analog-specific accumulation and particularly transformation of PBDEs, OH-PBDEs and MeO-PBDEs suggest that precautions should be taken when exploring phytoremediation strategy.

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Appendix A. Supplementary data

Supplementary data related to this article can be found at http://dx.doi.org/10.1016/j.envpol.2015.10.051.

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