Equilibrium Study, Modeling and Optimization of Model Drug Adsorption Process by Sunflower Seed Shells

Bahdja Hayoun 1,2,*, Mustapha Bourouina 1, Marta Pazos 2, Mª Angeles Sanromán 2 and Saliha Bourouina-Bacha 3

1 Department of Chemistry, Faculty of Exact Sciences, Campus Targa Ouzemmour, University of Bejaia, Bejaia 06000, Algeria; bouryas@yahoo.fr
2 CINTECX-Universidade de Vigo, Department of Chemical Engineering Campus As Lagoas-Marcosende, University of Vigo, 36310 Vigo, Spain; mcurras@uvigo.es (M.P.); sanroman@uvigo.es (M.A.S.)
3 Department of Process Engineering, Faculty of Technology, Campus Targa Ouzemmour, University of Bejaia, Bejaia 06000, Algeria; lgebej@yahoo.fr
* Correspondence: hayba.bahdja@gmail.com

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Abstract: The adsorption capacity of the medication methylthioninium chloride (MC) from aqueous solution onto sunflower seed shells (SSS), a low cost and abundant alternative adsorbent, was investigated in a batch system. The surface properties of the adsorbent were characterized by Fourier transform infrared (FTIR) spectroscopy, scanning electron microscopy (SEM), specific surface area (by using the Brunauer–Emmett–Teller equation), the liquid displacement method and pH\textsubscript{PZC}. The ability of SSS to remove the medication was assessed through kinetic, thermodynamic and equilibrium investigations. The adsorption efficiency of the SSS adsorbent for the removal of MC was evaluated considering the effects of its concentration, temperature, adsorption contact time, and the pH of the medium. The results obtained from the kinetic and isotherm studies show that the adsorption of the MC on SSS follows pseudo-second-order kinetics ($R^2 > 0.99$) and the Temkin isotherm model ($R^2 = 0.97$), respectively. The thermodynamic study showed that the adsorption was endothermic and spontaneous, according to its physisorption mechanism. The mathematical modeling of this process was carried out by using the surface response methodology of Box–Behenken. It was possible to deduce a statistically reliable regression equation that related the adsorption yield to the chosen operating parameters, that is, the initial MC concentration, the adsorbent dosage and the pH. Analysis of the variance indicated that the most influential parameters were the SSS dosage, the pH and their interaction and showed the optimal values for ensuring the best adsorption capacity of 95.58%.

Keywords: adsorption; experimental design; isotherm; methylthioninium chloride; kinetics; sunflower seed shells

1. Introduction

There has been a great deal of concern about the detection of pharmaceuticals and personal care products in soil, sediments, surface and groundwater [1] as they have become major pollutants due to their low biodegradability, high persistence, and facile bioaccumulation [2]. Among these products, methylthioninium chloride (MC), which is part of a type of medication derived from phenothiazines, has been widely used in a variety of medical treatments, such as methemoglobinemia [3] and lymphatic mapping/sentinel lymphadenectomy in staging melanoma and breast cancer [4]. It has also been used as a therapeutic agent in neurodegenerative diseases [5,6]. This product is not only used in the
pharmaceutical field, but it is also used in several other industrial fields including textiles, leather, cosmetics, paper, printing, plastic and food [7–9]. Therefore, MC is often present in the various effluents that are discharged into the environment. Thus, the continuous release of this pollutant into the environment significantly affects human health and aquatic systems [9,10].

Therefore, the treatment of these effluents by appropriate techniques is very necessary [9,11]. For this purpose, several physical, chemical and biological methods have been developed including coagulation and flocculation, membrane separation, reverse osmosis, oxidation and co-precipitation [9,12]. The use of an inexpensive and effective method is very important [11,13]. The adsorption technique has certain advantages; it is a simple and easy technique, it costs less and it is respectful of the environment [14,15]. Activated carbon is the most commonly used adsorbent due to its very high adsorption capacity [16–18], but the disadvantage is that it is expensive [16]. For this reason, several studies have been carried out to examine the use of low-cost adsorbents [19]. For example, various agricultural byproducts, clays [20], biomass [15], synthetic and natural polymers [21], have been evaluated. Thus, it is useful to find effective adsorbent materials, that are available locally and especially, low cost [7,22]. Indeed, sunflower seed shells, an agricultural byproduct of biomass, could be exploited in this area as they are a low-cost product, widely used and can be used without any treatment.

However, several factors greatly influence the adsorption process, such as the initial pollutant concentration, adsorbent dosage, pH and temperature of the solution, etc. These are important because they affect the adsorption process, particularly the adsorption capacity.

Accordingly, it is necessary to use methods based on statistical calculations like the Box–Behnken design (BBD), which is a useful tool because it is fast, efficient and economical in comparison to the classic univariate methods where only one factor is studied at a time. Indeed, the advantage of this technique is the collection of a large amount of information from a minimum of experimental trials, without sacrificing the accuracy of the results. The main aim of this study was to find the best operating conditions for the adsorption of MC on a low-cost adsorbent prepared from sunflower seed shells (SSS). Thus, a study was carried out to determine the influence of the operating parameters on the adsorption efficiency. The adsorption mechanism of the SSS is shown through the study of adsorption isotherms, kinetics and thermodynamics. The efficiency of the adsorption of SSS for the elimination of MC was optimized by evaluating the effects of three parameters, namely, the initial concentration of the drug, the dose of adsorbent and the pH of the solution. Modeling and optimization were carried out by the Box–Behnken design, which allowed us to determine the optimal values of these parameters and to ensure the high adsorption efficiency of MC on the SSS powder.

2. Materials and Methods

2.1. Reagents

Methylthioninium chloride (3,7-bis(dimethylamino)-phenazathionium chloride tetramethyle-thionine chloride), a model medication pollutant (CAS n°: 61-73-4) with a molecular weight of 373.9 g/mol, and maximum wavelength of 668 nm was purchased from Merck (Biochem Chemopharma). A range of concentrations from 20 to 150 mg/L were prepared by successive dilution of the stock solution (150 mg/L). Distilled water was used to prepare all solutions. The pH adjustment of the solutions was made by the addition of HCl (0.1 M) or NaOH (0.1 M) using a pH-meter (Boeco).

2.2. Preparation of Adsorbent

The sunflower seed shells (SSS) was purchased from a local store (Béjaia, Algeria). First, they were washed several times with distilled water and then dried at 60 °C for one day. Then, the dry product was crushed and then sieved to obtain a particle size less than or equal to 0.3 mm. After that, it was stored in closed bottles.
2.3. Characterization of Adsorbent

To analyze the functional groups, a Fourier transform infrared (FTIR) spectroscopy study was carried out using a 630 Cary spectrometer using a small amount of adsorbent before and after adsorption of MC.

A scanning electron microscope (SEM) FEI Quanta FEG 250 from the University of Lyon 1, France, was used to study the morphology of the adsorbent.

The specific surface area was determined by the adsorption of $N_2$ at 77.3 K using a surface analyzer Nova 2000e (quantachrome instruments) using the BET (Brunauer–Emmett–Teller) equation model. In addition, the total pore volume and average pore diameter were measured.

The real density, apparent density (bulk density), porous volume and porosity of the adsorbent were measured by the liquid displacement method. The real density $\rho_{\text{real}}$ ($\text{g/cm}^3$) was determined by filling a pycnometer with a well-defined mass of the adsorbent note, $m_{\text{ads}}$ and then the pycnometer was adjusted to its volume ($V_{\text{pyc}}$) by adding a well-defined volume of methanol (99% purity) and was weighed. The calculation was made by using the following equation [23].

$$\rho_{\text{real}} = \frac{m_{\text{ads}}}{V_{\text{real of adsorbent}}} \quad (1)$$

However, the apparent density $\rho_{\text{app}}$ ($\text{g/cm}^3$) (bulk density) was determined using a calibrated cylinder, which was filled with a given adsorbent weight ($m_{\text{ads}}$) and then the cylinder was tapped until a minimum volume was recorded. This represents the apparent volume ($V_{\text{app}}$) of the adsorbent which was evaluated using the following equation [23].

$$\rho_{\text{app}} = \frac{m_{\text{ads}}}{V_{\text{app}}} \quad (2)$$

The porous volume $P.V$ ($\text{cm}^3/\text{g}$) and porosity $\varepsilon$ of the adsorbent were calculated using the results of the densities as well as the following equations [23].

$$P.V = \frac{1}{\rho_{\text{app}}} - \frac{1}{\rho_{\text{real}}} \quad (3)$$

$$\varepsilon = 1 - \frac{\rho_{\text{app}}}{\rho_{\text{real}}} \quad (4)$$

The determination of the zero point charge of SSS (ZPC) was performed by the salt addition method [24]. A 0.05 M NaCl solution was prepared, transferred to a series of 50 mL beakers and the pH of each solution was adjusted from 2 to 12 by adding 0.1 M HCl or 0.1 M NaOH solution. The initial pH of these solutions (noted as $pH_i$) was measured. Then, 0.15 g of SSS was added to each of the beakers, which were tightly closed using parafilm paper and then stirred at 300 rpm at room temperature. The final pH value was measured after 48 h of agitation, the noted $pH_f$ and the ZPC value of the surface was calculated from the plot of $\Delta pH= f(pH_f)$.

2.4. Adsorption Experiments

Adsorption experiments were carried out by agitating the SSS powder in an Erlenmeyer flask (500 mL) that contained 250 mL of the MC solution of desired concentration and pH in a magnetic stirrer at 300 rpm. The experiments were carried out with varying adsorbent dosage (1–3 g/L), initial MC concentration (20 –150 mg/L), temperature (25–40 °C) and pH (2–12). The samples were withdrawn from the Erlenmeyer at predetermined time intervals. The MC solution was separated from the adsorbent using a Hettich brand centrifuge at 5000 rpm for 3 min and analyzed for their MC content spectrophotometrically at the appropriate wavelength ($\lambda_{\text{max}} = 264$ nm) using a UV spectrophotometer (Shimadzu). The effect of pH was studied by adjusting the pH of the MC solutions using HCl and
NaOH. All experiments were performed in duplicate, all data was calculated and average values are taken to represent the results.

The uptake of MC adsorbed onto SSS per unit mass was calculated using the following equation:

\[ q_t = \frac{(C_0 - C_t)V}{m} \]  

where \( q_t \) is the amount of MC adsorbed onto SSS (mg/g) at time \( t \), \( C_0 \) and \( C_t \) are the liquid phase concentrations of MC at initial and time \( t \), respectively. \( V \) (L) is the volume of MC solution and \( m \) (g) is the mass of the SSS adsorbent used. The MC removal efficiency was calculated as follows:

\[ Y\% = \frac{(C_0 - C_t)}{C_0} \times 100 \]  

and the average relative error (ARE) was calculated as follows [25]:

\[ ARE = \sum_{i=1}^{N} \frac{|q_{e, \text{exp}} - q_{e, \text{model}}|}{q_{e, \text{exp}}} \]  

where \( q_{e, \text{exp}} \) and \( q_{e, \text{model}} \) (mg/g) are the uptake at equilibrium measured (exp) and calculated by the isotherm model (model), respectively. We also used the same relationship to determine the mean relative errors for the kinetic models.

2.5. Data Analysis

2.5.1. Kinetics Study

The study of the adsorption kinetics helps to determine the adsorption equilibrium time as a function of the adsorbent used and the operating conditions. It also allows us to understand the phenomena at the interface between the absorbing material and the pollutant molecules in terms of the diffusion and adsorption mechanism [9,18]. Three kinetic models: the pseudo-first order, the pseudo-second order and the intra-particle diffusion model [26–28] were used to analyze the kinetic data of MC adsorption on SSS. These models can be written as follows:

pseudo-first order model:

\[ q_t = q_1(1 - e^{-k_1t}) \]  

pseudo-second order model:

\[ q_t = \frac{q_2^2k_2t}{1+q_2k_2t} \]  

intra-particle diffusion model:

\[ q_t = K_d \sqrt{t} + c \]  

where \( q_1 \) and \( q_2 \) are the theoretical uptakes at equilibrium (mg/g), \( q_t \) (mg/g) is the uptake at time \( t \), and \( k_1 \) (1/min) and \( k_2 \) (g/mg min) are the pseudo-first-order rate constant and the pseudo-second-order rate constant, respectively. \( K_d \) (mg.g.-1/√min) is the intra-particle diffusion constant and \( c \) (mg/g) is a constant related to the thickness of the boundary layer.

2.5.2. Equilibrium Study

The adsorption isotherms can be defined as the set of adsorption equilibrium states, at a given temperature. Several models have been used in the literature to describe the experimental data of adsorption isotherms [7]. In this study, we chose three models [29–31] and we calculated the various parameters by the equations defined as follows:
Langmuir model:
\[ q_e = \frac{q_m K_L C_e}{1 + K_L C_e} \]  
(11)

Freundlich model:
\[ q_e = K_F C_e^{1/n_F} \]  
(12)

Temkin model:
\[ q_e = B_T \ln(K_T C_e) \]  
(13)

where \( q_e \) and \( C_e \) are the uptake and concentration at the equilibrium, respectively. \( q_m \) (mg/g) is the maximum uptake of the adsorbent, \( K_L \) (L/mg) is the Langmuir constant related to the free energy of adsorption. \( K_F \) (mg/g) (L/g)^n and \( n_f \) are Freundlich constants, which are indicators of adsorption capacity and adsorption intensity, respectively. \( B_T \) (KJ/mol) and \( K_T \) (L/g) are the Temkin constants.

2.5.3. Thermodynamic Study

Temperature is an indicator of the nature of the adsorption, that is, whether it is an exothermic \((\Delta H^o < 0)\) or endothermic \((\Delta H^o > 0)\) process. Measurement of the adsorption heat \(\Delta H^o\) is the main criterion for differentiating chemisorption from physisorption [32].

\(\Delta G^o\), \(\Delta H^o\) and \(\Delta S^o\) are the thermodynamic parameters that determine the feasibility and spontaneity of the adsorption process. They were determined using the following relations [7,33]:

\[ K_{ad} = \frac{q_e}{C_e} \]  
(14)

\[ \Delta G = -RT \ln K_{ad} \]  
(15)

\[ \ln K_{ad} = \frac{\Delta S}{R} - \frac{\Delta H}{RT} \]  
(16)

where, \( K_{ad} \) (L/g) is the distribution constant, \( q_e \) (mg/g) represents the equilibrium solid-phase concentration, \(\Delta G^o\) (J/mol) is the standard free energy change, \(\Delta H^o\) (J/mol) is the standard enthalpy change and \(\Delta S^o\) (J/mol.K) is the standard entropy change, \( R \) is the universal gas constant (J/mol. K), and \( T \) is the absolute temperature (K).

2.6. Experimental Design and Statistical Analysis

A three-level, three-factor, 15-assay Box–Behnken factorial design was used to optimize MC adsorption on the adsorbent prepared from SSS using the software Minitab (version 17). The advantages of BBD are that it avoids having to experiment in extreme conditions and it does not involve combinations for which all factors are simultaneously at their highest or lowest points [19,34,35].

The independent variables selected for the study are the initial concentration of the aqueous solution MC \((X_1)\), the adsorbent dosage \((X_2)\) and the solution pH \((X_3)\), which are considered the factors that have the most influence on the adsorption of the model medication. However, the adsorbent-adsorbate contact time is set at 90 min, as this is the time needed to reach equilibrium. These independent variables were studied at three different levels, low (−1), middle (0) and high (+1), as shown in Table 1 [19]. The predicted response \( Y \) is related to these factors through a quadratic polynomial as follows:

\[ Y% = a_0 + a_1 X_1 + a_2 X_2 + a_3 X_3 + a_{11} X_1^2 + a_{22} X_2^2 + a_{33} X_3^2 + a_{12} X_1 X_2 + a_{13} X_1 X_3 + a_{23} X_2 X_3 \]  
(17)

where \( Y \) is the measured response, \( a_0 \) is the intercept, \( a_1, a_2, a_3 \) are linear coefficients, \( a_{ii=1,3} \) are quadratic coefficients, and \( a_{ij=1,3,i,j} \) are interactive coefficients.
Table 1. Range of variation of operating parameters.

| Symbol | Variable                | Level (−1) | Level (0) | Level (+1) |
|--------|-------------------------|------------|-----------|------------|
| $X_1$  | $C_i$ (mg/l) MC initial concentration | 20         | 55        | 90         |
| $X_2$  | m (g/L) SSS dosage      | 1          | 2         | 3          |
| $X_3$  | pH                      | 4          | 7         | 10         |

3. Results

3.1. Characterization of SSS Adsorbent

3.1.1. Fourier Transform Infrared Spectroscopy

The FTIR spectra of the adsorbent before and after adsorption of MC are presented in Figure 1. The FTIR spectra of the SSS clearly show several absorption peaks which reflect the complex nature of the material examined. The broad absorption peak around 3331.5 cm$^{-1}$ was indicative of the existence of bonded hydroxyl groups on the surface of SSS. The peaks around 2912 and 2102 cm$^{-1}$ were assigned to the stretching vibration and bending vibration of the C–H bond, respectively [7,16,36]. The broadband at 1732 cm$^{-1}$ was assigned to the carbonyl group (C=O). The peaks associated with the stretching vibration in aromatic rings (C=O and –NH of amide groups) were observed at 1591 and 1505 cm$^{-1}$. The weak peaks around 1419 cm$^{-1}$, 1369 cm$^{-1}$ and 1316 cm$^{-1}$ were attributed to the vibration of the –CN bond [9,16]. The peak at nearly 1234 cm$^{-1}$ was assigned to the bending vibration of carboxylic groups while deformation related to C–O bond was observed at 1040 cm$^{-1}$ [32].

![Figure 1. FTIR spectra of sunflower seed shells (SSS) before and after adsorption of methylthioninium chloride (MC).](image-url)

The strong peak at 1029 cm$^{-1}$ was attributed to the Si (Si–O) valence vibrational binding. However, the weak peaks that intervene in the range 450–650 cm$^{-1}$ can be attributed to the Si-reflection vibration binding S–O. All these peaks indicate the massive presence of oxygenated groups on the surface of SSS [36]. However, when the two spectra are compared, it is clear that the SSS spectrum after adsorption of MC shows a slight difference, namely, a generally slight shift in all peaks as well as an increase in the intensity of the peaks around 1594, 1316 and 983 cm$^{-1}$. These changes are attributed to the loading of the SSS surface with the molecule of the model drug MC [22].
3.1.2. Scanning Electro-Microscopy (SEM)

The morphology of SSS examined by SEM, is illustrated in Figure 2. The SEM micrograph clearly shows that the SSS has an irregular and porous surface and the pores form a kind of honeycomb. This structure allows the MC solution to penetrate into the pores and attach to host sites. Indeed, this ability of the SSS to adsorb the MC was confirmed visually by the staining of the adsorbent after the adsorption test.

Figure 2. SEM micrograph of SSS.

3.1.3. BET Specific Surface Area and Other Characteristics of SSS

The physico-chemical characteristics of SSS are summarized in Table 2. As observed, the BET surface area of the SSS was found to be 6.106 m²/g, with a total pore volume of 9.689 cm³/g, and average pore diameter of 3.17330 Å; this shows that the SSS have better surface characteristics as a natural adsorbent, without any treatment.

In addition, the SSS had a very high porosity of about 90% which is in agreement with the SEM characterization, as shown in Figure 2. On the other hand, the real and apparent density, as well as the zero load point were determined. These gave us the same resolution as the other parameters. Guediri et al. [37] showed a porosity of 79% with a bulk density of 0.026 g/cm³ for orange peels treated with phosphoric acid (OP–H₃ PO₄). Stavrinou et al. [38] determined BET surface areas of 0.8549 m²/g and 0.6140 m²/g for cucumber peel and potato peel, respectively.
3.2. Adsorption Equilibria

3.2.1. Effect of Adsorbent–Adsorbate Contact Time and Initial MC Concentration

The effect of the contact time (0–120 min) and the initial MC concentration (20, 30, 60, 90, 120, 150 mg/L) on the adsorption efficiency are shown in Figure 3. The adsorption efficiency of MC increases with the contact time and reaches equilibrium at about 90 min with a maximum removal value of 90.07% for an initial concentration of 20 mg/L, pH = 6 of the solution and 2 g/L of the SSS dosage. The rapid increase in MC adsorption efficiency in the first 30 min can be attributed to the availability of the uncovered surface as well as the high number of empty active sites at the initial time of the process on the adsorbent surface. It is clear that the elimination efficiency decreases from 90.07 to 69.77% when the initial concentration of MC decreases from 20 to 150 mg/L. This result can be explained by considering that at a lower initial MC concentration, the ratio of the adsorbent active sites to the total adsorbate is high and therefore all adsorbate molecules can interact with the adsorbent and be removed from the solution [7, 17]. However, the uptake of MC at equilibrium state climbed from 8.96 mg/g to 52.50 mg/g, when the initial drug concentration increased from 20 to 150 mg/L. Indeed, it can be assumed that because of the increase in the content of MC in solution, and therefore the increase in the concentration gradient, the arrival of the molecules of the adsorbate towards the surface of the adsorbent becomes easier, which induces the higher occupation of active sites [18–39]. Therefore, 60 mg/L was selected as the optimum initial concentration of MC for the adsorption experiments.

![Figure 3](image_url)

**Figure 3.** Evolution of adsorption efficiency (a) and uptake (b) of MC at equilibrium in terms of contact time and initial MC concentration (conditions: SSS dose: 2 g/L, pH 6.0, flask shaking at 300 rpm at 25 °C).

3.2.2. Effect of Adsorbent Dosage

The dependence of the MC adsorption efficiency on the dosage of SSS (1–3 g/L) for an initial concentration of 60 mg/L, a solution pH 6 and a contact time of 120 min is illustrated in Figure 4. The adsorption rate increases with the dose of SSS and stabilizes from 2.5 g/L. A significant improvement in the adsorption of MC from 66.39% to 89.04% is observed when the SSS dosage changes from 1 to 2.5 g/L. This result is due to the increase in functional groups and adsorption sites, which is linked to the properties of the adsorbent [7, 17].

On the other hand, the increase in sorbent dose from 1 g/L to 3 g/L caused a decrease in the uptake of MC from 39.72 mg/g at 17.60 mg/g, respectively, which is entirely logical given that “q” is inversely proportional to the total quantity of adsorbent in suspension [39].
positive ions in solution. Consequently, the electrostatic attraction of the cationic pollutant decreases pH below pH \( \text{pH}_{ZPC} \) and its adsorption becomes weaker. As the pH increases, the concentration of H ions decreases and its adsorption efficiency may also be due to the protonation of functional groups on the adsorbent surface. As already shown, the pH \( \text{pH}_{ZPC} \) of SSS is 7.14; therefore, the surface of the adsorbent will be positively charged at a pH below \( \text{pH}_{ZPC} \) and negatively at a pH above \( \text{pH}_{ZPC} \). Thus, adsorption of MC (cationic substance) is favorable in basic medium. On the other hand, at acidic pHs, the surface of the adsorbate is excessively protonated, thus, the overall load of the adsorbent is positive. The cationic molecule MC releases positive ions in solution. Consequently, the electrostatic attraction of the cationic pollutant decreases and its adsorption becomes weaker. As the pH increases, the concentration of H ions decreases and more MC molecules occupy the active sites of the SSS until saturation [7,8].

3.2.3. Effect of pH

The initial pH of the adsorption medium is one of the most influential parameters in the process of adsorption, it affects both the surface charges of the adsorbent and the degree of ionization of the adsorbate. In the present study, the adsorption experiments were performed in the initial solution pH range from 2 to 12 (Figure 5). The results show a significant increase in the adsorption efficiency and uptake from 19.83% and 4.42 mg/g to 88.33% and 21.26 mg/g when the pH values increased from 2 to 4, respectively, followed by a slight increase for pH 12 (95.97%, 22.75 mg/g). The increase in adsorption efficiency may also be due to the protonation of functional groups on the adsorbent surface. As already shown, the pH \( \text{pH}_{ZPC} \) of SSS is 7.14; therefore, the surface of the adsorbent will be positively charged at a pH below \( \text{pH}_{ZPC} \) and negatively at a pH above \( \text{pH}_{ZPC} \). Thus, adsorption of MC (cationic substance) is favorable in basic medium. On the other hand, at acidic pHs, the surface of the SSS is excessively protonated, thus, the overall load of the adsorbent is positive. The cationic molecule MC releases positive ions in solution. Consequently, the electrostatic attraction of the cationic pollutant decreases and its adsorption becomes weaker. As the pH increases, the concentration of H ions decreases and more MC molecules occupy the active sites of the SSS until saturation [7,8].

![Figure 4](image1.png)

**Figure 4.** Effect of adsorbent dose on the adsorption efficiency (a) and uptake (b) of MC onto SSS (conditions: initial MC concentration 60 mg/L, pH 6, flask shaking at 300 rpm at 25 °C).

![Figure 5](image2.png)

**Figure 5.** Effect of pH on the adsorption efficiency (a) and uptake (b) of MC onto SSS (conditions: initial MC concentration 60 mg/L, adsorbent dose: 2.5 g/L, flask shaking at 300 rpm at 25 °C).
3.2.4. Effect of Temperature

One of the important parameters affecting adsorption capacity is the temperature of the solution. To determine the effect of temperature on the adsorption of MC onto SSS, a study was carried out in the range of 25–40 °C, and the results are shown in Figure 6. The obtained results reveal a slight increase in the MC adsorption efficiency, from 89.04% at 25 °C to 91.97% at 40 °C, and an uptake of 21.41 mg/g at 25 °C to 22.07 mg/g at 40 °C with the increase in temperature. This behavior was evident because as the temperature increases, the viscosity of the solution decreases, which is favorable for the adsorption steps of external transfer, and adsorbate diffusion inside the adsorbent solid. This increase may be due to the increased mobility of MC, which allows it to penetrate the pores of the sample, the increased chemical interactions between the adsorbate and the adsorbent surface functions, or changes in chemical potentials that correlate with the solubility of the adsorbate species. This small variation is indicative of physisorption and suggests that the adsorption is endothermic [32,33].

![Figure 6. Effect of temperature on the adsorption efficiency (a) and uptake (b) of MC onto SSS](image)

3.3. Kinetics of Adsorption

The adsorption kinetic is an essential tool in the study of the adsorption mechanism. Therefore, pseudo-first order, pseudo-second order, and intra-particle diffusion models were used to describe the adsorption process. The constants obtained from these three kinetic models are given in Table 3. The results show that the pseudo-second-order model is better adapted because of the high correlation coefficient (R² > 0.99), good agreement between equilibrium uptakes calculated by the model (qₑ) and the experimentally obtained (qₑ), and the weak average relative errors. Similar results have been observed in the adsorption of MC on activated carbon [40] and factory-rejected tea [18]. Adsorption of MC onto the adsorbent may comprise several steps, but in this study, kinetic models of pseudo-first order and pseudo-second order could not determine the adsorption mechanism. Hence, intra-particle scattering was used to clarify the adsorption mechanism. The results of this study also showed that adsorption of MC takes place in three stages. The first step is attributed to the transfer of MC in the solution to the liquid phase located near the surface of the adsorbent. The second is the diffusion of the adsorbate through the liquid film surrounding the adsorbent grain while the third corresponds to the diffusion of the MC in the pores of the adsorbent. This last step is the limiting step; it is the slowest step of the process [9,41].
Table 3. Kinetic parameters of the adsorption of MC on the SSS following different kinetic models.

| Initial Concentration of MC (mg/L) | 20   | 30   | 60   | 90   | 120  | 150  |
|-----------------------------------|------|------|------|------|------|------|
| q_e (exp.) (mg/g)                 | 9.029| 13.927| 25.893| 37.389| 48.51| 51.522|

Pseudo-first order model

| q_1 (mg/g)        | 0.457 | 1.528 | 7.649 | 12.222| 12.958| 15.979|
|-------------------|-------|-------|-------|-------|-------|-------|
| k_1 (1/min)       | 0.0358| 0.0646| 0.0744| 0.0472| 0.0371| 0.038 |
| R^2               | 0.87  | 0.944 | 0.969 | 0.965 | 0.989 | 0.972 |
| ARE (%)           | 97.191| 92.531| 79.186| 78.708| 84.380| 82.813|

Pseudo-second order model

| q_2 (mg/g)        | 9.047 | 12.453| 26.247| 38.759| 44.444| 54.054|
|-------------------|-------|-------|-------|-------|-------|-------|
| k_2 (g/mg min)    | 0.243 | 0.016 | 0.027 | 0.009 | 0.003 | 0.005 |
| R^2               | 1     | 0.994 | 0.999 | 0.999 | 0.991 | 0.999 |
| ARE (%)           | 1.455 | 28.099| 2.424 | 3.064 | 26.584| 7.421 |

Intra-particle diffusion model

| K_d1 (mg/g \(\sqrt{\text{min}}\)) | 0.338 | 1.137 | 1.167 | 2.62  | 2.26  | 2.62  |
|-----------------------------------|-------|-------|-------|-------|-------|-------|
| C_1 (mg/g)                        | 7.863 | 9.858 | 13.433| 20.461| 31.975| 23.101|
| R^2                               | 1     | 0.993 | 0.874 | 0.914 | 0.972 | 0.978 |
| K_d2 (mg/g \(\sqrt{\text{min}}\)) | 0.048 | 0.071 | 0.08  | 0.369 | 0.848 | 0.963 |
| C_2 (mg/g)                        | 8.58  | 13.273| 25.121| 33.85 | 40.324| 42.743|
| R^2                               | 0.875 | 0.889 | 0.741 | 0.977 | 0.998 | 0.806 |
| ARE (%)                           | 3.743 | 8.439 | 15.498| 19.304| 3.286 | 17.329|

3.4. Adsorption Isotherms

To carry out a more detailed study of the adsorption mechanism, the adsorption isotherms were adjusted to the Langmuir, Freundlich and Temkin models. The constants obtained from the equations of these three models are summarized in Table 4. The results show that the Temkin model is the one that best simulates this process because of the higher correlation coefficient \(R^2 > 0.970\), and a lower mean relative error (ARE = 9.804%). Similar results were observed in the study of adsorption of MC via maize silk powder [25]. For the Langmuir model, the value of the separation factor \(R_L\) describes the feasibility of the adsorption process. In our study, \(R_L\) values vary between 0.071 and 0.652 indicating favorability of adsorption [42]. The latter also allowed us to determine the maximum uptake of MC on SSS, \(q_m = 63.29 \text{ mg/g} \) at equilibrium time \(t_{equ} = 90 \text{ min}\). As a comparison, Ogununtimein [39] determined a maximum uptake of \(q_m = 0.0365 \text{ mg/g} \) for the adsorption of MC onto SSS treated with sodium hydroxide (0.05 M NaOH) at \(t_{equ} = 120 \text{ min}\), and El-Halwany [43] determined a maximum uptake of \(q_m = 39.4 \text{ mg/g} \) for the adsorption of MC onto activated carbon prepared by chemical activation of SSS uses phosphoric acid (\(\text{H}_3\text{PO}_4\)) at \(t_{equ} = 150 \text{ min}\).

The Freundlich isotherm model allowed us to determine the parameter \(n_F\), which determines the adsorption effectiveness. According to the results, the calculated value of the Freundlich isotherm model constant, \(n_F\) is greater than 1 (\(n_F = 1.879\)), which confirms that the adsorption phenomenon is favorable [42].
Table 4. Parameters of the adsorption isotherm models.

| Isotherms Models | Parameters | Values  | R²   | ARE (%) |
|------------------|------------|---------|------|---------|
| Langmuir         | q_m (mg/g) | 63.29   | 0.986| 57.611  |
|                  | K_L (L/mg) | 0.275   |      |         |
|                  | R_L        | 0.071–0.652 |      |         |
| Freundlich       | K_F (mg/g) (L/g)^n | 8.005 | 0.945| 14.365  |
|                  | n_F        | 1.872   |      |         |
| Temkin           | K_T (L/g)  | 1.093   | 0.970| 9.804   |
|                  | RT/b (B_T(KJ/mol)) | 13.508 |      |         |

3.5. Thermodynamic Study

Since temperature plays an important role in MC adsorption, the thermodynamic adsorption parameters were calculated to evaluate the thermodynamic feasibility and the spontaneous nature of the process (Table 5). The results showed that the adsorption of MC on SSS was spontaneous and favorable at all experimental temperatures as a result of the negative values of ΔG; furthermore, the ΔG values decrease with increasing temperature (from −2.967 KJ/mol for 25 °C to −3.953 KJ/mol for 40 °C), which indicates that this process becomes less favorable as the temperature increases. Similar results were observed by Gao et al. [42] in the study of adsorption of MC onto activated carbon produced from tea (Camellia sinensis L.) seed shells. The positive value of ΔH (ΔH = 58.461 KJ/mol) suggests that the process is endothermic [33]. We can see that the entropy value ΔS is positive, almost constant and weak for the whole range of temperatures studied (0.206–0.199 KJ/mol K); this indicates an increase in the randomness of interactions in the diffusion layer during the adsorption process [17].

Table 5. Obtained thermodynamic parameters in the range of tested temperatures.

| T (°K) | ΔG° (KJ/mol) | ΔS° (KJ/mol K) | ΔH° (KJ/mol) |
|--------|--------------|----------------|-------------|
| 298    | −2.967       | 0.206          |             |
| 303    | −3.245       | 0.204          |             |
| 308    | −3.314       | 0.201          |             |
| 313    | −3.953       | 0.199          |             |
|        |              |                | 58.461      |

3.6. Optimization and Modeling by Response Surface

In this study, a three-level, three-factor, 15-assay BBD was used to optimize MC adsorption on adsorbent prepared from SSS. The independent variables evaluated for the study are the initial concentration of the aqueous solution MC (X_1), the adsorbent dosage (X_2) and the solution pH (X_3) (Table 1). The experiments were performed to optimize the adsorption efficiency of the medication on the prepared adsorbent. The obtained results are summarized in Table 6, in which the response evaluated was the MC removal (Y).

The obtained responses were analyzed using Minitab 17 software. The polynomial equation that describes the main effects, the interaction effects and the quadratic effects of the selected independent variables show a significant influence on the responses [44]. The results of the analysis were evaluated using ANOVA variance analysis and are presented in Table 7.
Table 6. Matrix of experiments with coded, uncoded factors and the response.

| N0 | X1 | X2 | X3 | C_i (mg/L) | m (g/L) | pH  | Y (%)     |
|----|----|----|----|------------|---------|-----|-----------|
| 1  | -1 | -1 | 0  | 20         | 1       | 7   | 88.5539   |
| 2  |  1 | -1 | 0  | 90         | 1       | 7   | 75.1651   |
| 3  | -1 |  1 | 0  | 20         | 3       | 7   | 96.5097   |
| 4  |  1 |  1 | 0  | 90         | 3       | 7   | 93.3073   |
| 5  | -1 |  0 | -1 | 20         | 2       | 4   | 77.9135   |
| 6  |  1 |  0 | -1 | 90         | 2       | 4   | 77.3905   |
| 7  | -1 |  0 |  1 | 20         | 2       | 4   | 96.6233   |
| 8  |  1 |  0 |  1 | 90         | 2       | 4   | 92.3212   |
| 9  |  0 | -1 | -1 | 55         | 1       | 4   | 59.7206   |
| 10 |  0 |  1 | -1 | 55         | 3       | 4   | 89.1495   |
| 11 |  0 | -1 |  1 | 55         | 1       | 10  | 85.9411   |
| 12 |  0 |  1 |  1 | 55         | 3       | 10  | 96.7700   |
| 13 |  0 |  0 |  0 | 55         | 2       | 7   | 91.1453   |
| 14 |  0 |  0 |  0 | 55         | 2       | 7   | 92.5922   |
| 15 |  0 |  0 |  0 | 55         | 2       | 7   | 93.4458   |

Table 7. Analysis of variance (ANOVA) of the model.

| Source           | DF | Adj SS   | Adj MS  | F-Value | p-Value |
|------------------|----|----------|---------|---------|---------|
| Model            | 9  | 1458.65  | 162.072 | 17.98   | 0.003   |
| Linear           | 3  | 1176.94  | 392.312 | 43.52   | 0.001   |
| C_i              | 1  | 57.33    | 57.332  | 6.36    | 0.053   |
| m                | 1  | 550.38   | 550.385 | 61.05   | 0.001   |
| pH               | 1  | 569.22   | 569.219 | 63.14   | 0.001   |
| Square           | 3  | 165.71   | 55.238  | 6.13    | 0.040   |
| C_i·C_i          | 1  | 0.66     | 0.657   | 0.07    | 0.798   |
| m·m              | 1  | 47.55    | 47.550  | 5.27    | 0.070   |
| pH·pH            | 1  | 128.99   | 128.986 | 14.31   | 0.013   |
| 2-way Interaction| 3  | 116.00   | 38.667  | 4.29    | 0.075   |
| C_i·m            | 1  | 25.94    | 25.940  | 2.88    | 0.151   |
| C_i·pH           | 1  | 3.57     | 3.571   | 0.40    | 0.557   |
| m·pH             | 1  | 86.49    | 86.490  | 9.59    | 0.027   |
| Error            | 5  | 45.08    | 9.015   |         |         |
| Lack-of-fit      | 3  | 42.37    | 14.124  | 10.44   | 0.089   |
| Pure error       | 2  | 2.70     | 1.352   |         |         |
| Total            | 14 | 1503.73  |         |         |         |

The high values of the correlation coefficient (R²) and the adjusted correlation coefficient (R²-Adj), 0.97 and 0.91, respectively, show that the proposed model explains more than 91% of the variations in the experimental response [35].

Similar results were determined by Gulen and Demircivi [45] who determined values of 0.9224 and 0.8227 for R² and R²-Adj coefficients, respectively. So, the authors concluded that these values indicated good fitness and high significance for the statistical model. Elmouarki et al. [46] also reported that the obtained R² coefficients are in reasonable agreement with the R²-Adj and more than 95% of the response can be well predicted by the models.

The quadratic model is written in the following form:

\[
Y\% = -4.4 - 0.121 \, C_i + 29.50 \, m + 15.60 \, pH - 0.67 \, pH^2 - 1.550 \, m \cdot pH
\]  

(18)
The quadratic model well describes the adsorption efficiency (Y%) of MC on the adsorbent (SSS). It is significant and adequate as proved by the p-value (p-value = 0.003) with only 3% of total unapplied variations for the model (R² = 97%) [35,44].

Figure 7 shows the 3-D response surfaces of the individual and interaction effects of the parameters that influence the MC adsorption process on SSS. The 3-D curves were plotted by varying two parameters in the experimental domain while the third parameter is kept constant (at the center level). It was found that high adsorption efficiency can be attained at low initial concentrations and high pH (Figure 7b). With respect to the effect of amount on adsorption efficiency, it can be seen in Figure 7a,c that high efficiency can be obtained when the adsorbent dosage of SSS is high [35]. This is confirmed by Figure 8 which shows the optimal conditions for maximum MC adsorption efficiency onto SSS.

![Figure 7](image1.png)

**Figure 7.** Response surface plots of Y% or R%: (a) interaction Ci×m, (b) interaction Ci×pH, (c) interaction m×pH.

![Figure 8](image2.png)

**Figure 8.** Optimal conditions for maximum MC adsorption efficiency onto SSS: (a) Ci (mg/L) MC initial concentration, (b) m (g/L) SSS dosage and (c) pH.
Model terms with a \( p \)-value < 0.05 are significant at a 95% confidence level, while terms with a \( p \)-value < 0.05 do not affect the response [44]. Based on the analysis of the statistical results, it can be concluded that the resulting quadratic model is adequate and appropriate for describing the relationship between experimental parameters and adsorption efficiency [35].

Moreover, the optimization of the operating conditions was conducted using the maximum response function \( (Y = 98.77\%) \) which could be obtained by taking the following operating conditions: \( C_i = 20 \text{ mg/L}, \ m = 2.4 \text{ g/L} \) and \( \text{pH} = 8.9 \).

The predicted optimal adsorption yield was verified by performing three tests under optimal operating conditions \( (Y = 95.58\%) \). The standard deviation of the achieved experimental results in comparison to the theoretical results is 2.25. Thus, it can be concluded that the model describing the adsorption of MC on SSS is valid.

4. Conclusions

In this work, two approaches were used, the first was the evaluation of MC adsorption on SSS by kinetic, thermodynamic and equilibrium studies; and the second was modeling with an experimental Box–Behnken design. The equilibrium data indicated that the adsorption was fast and achieved within 120 min of contact time. Analyses of the kinetic and isotherm models revealed that the experimental data were well fitted by the pseudo-second order model and the Temkin isotherm, respectively. The adsorption process was shown to be controlled by intra-particle diffusion in the later stages. The thermodynamic study showed that adsorption was endothermic, spontaneous, and thus followed a physisorption mechanism. Furthermore, BBD design was appropriate for determining the optimal conditions for MC adsorption onto SSS. The optimal conditions of adsorption are an adsorbent dose of 2.4 g/L, an initial MC concentration of 20 mg/L and \( \text{pH} = 8.9 \) with a solution temperature of 25 °C; in these conditions the adsorption efficiency was 95.58%. Thus, SSS may be considered as a cheap and excellent adsorbent that shows excellent adsorptive characteristics for the removal of MC from aqueous solutions.

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