Two-phased Mass Rarity and Extinction in Land Plants During the End-Triassic Climate Crisis

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Greenhouse gas emissions from large-scale volcanism in the Central Atlantic Magmatic Province is considered to have caused the end-Triassic mass extinction (201.5 million years ago), but the impact on land plants has been debated. Here, abundance changes in spores and pollen record the devastating effects this volcanic induced climate crisis had on coastal and near-coastal lowland mire vegetation around the European epicontinental sea and the European Tethys margin. Combined stress from rising air temperatures and changing climate at the onset of the crisis was exacerbated by a rapidly rising sea-level resulting in fragmentation and destruction of coastal and near-coastal lowland mire habitats, causing mass rarity and extinctions primarily in gymnosperm trees and shrubs adapted to these environments. The devastation of these habitats was further amplified by a subsequent sea-level fall leaving pioneering opportunists and herbaceous survivors to colonize disturbed areas in an environment stressed by increased wildfire activity and enhanced soil erosion. The pioneering flora was severely decimated in a second mass rarity phase and ultimately extirpated. The second mass rarity phase occurred just prior to and at the onset of a prominent negative excursion in δ¹³Corg. A subsequent sea-level rise appears to have restored some of the near-coastal mire habitats allowing some of plants to recover. The supraregional mass rarity during the end-Triassic crisis affected both previously dominant as well as rare plants and this resonates with ongoing and future climate change and attests to the vulnerability of coastal and lowland vegetation, especially rare plant species, to climatic and environmental disturbances, where rising sea-level threatens entire ecosystems.

Keywords: palynology, palaeoclimate, global warming, sea-level, mass extinction, Triassic—Jurassic boundary, spores, pollen

1 INTRODUCTION

The end-Triassic mass extinction is generally recognized as one of the five major mass extinctions of the Phanerozoic (Sepkoski, 1996; McGhee et al., 2013), but the severity of the crisis on the vegetation is still debated. Amongst the land plants only seed ferns of the family Peltaspermaceae became extinct during the end-Triassic crisis (McElwain and Punyasena, 2007). However, on lower taxonomic levels extinctions are estimated to be higher, with up to 95% of species locally (McElwain et al., 1999). Estimating the severity of the crisis in land plants is complicated by provinciality and conflicting records between extinctions based on fossil leaf taxa and those based on spores and pollen (Bond and Wignall, 2014; Lindström, 2016; Barbacka et al., 2017). However, extinctions in spore and pollen taxa...
Triassic was profound, with replacement or major disruptions of perspective, the crisis amongst land plants at the end of the sites with little potential for fast burial. From an ecological macroscopic remains, e.g. epiphytic plants or plants growing in or in the earliest Hettangian were plants with ecological producing plants that disappeared at the end of the Rhaetian unknown (Lindström, 2016). Likely, many of the spore-went extinct, the parent plant affinity is not fully resolved or even unknown (Lindström, 2016). Likely, many of the spore-producing plants that disappeared at the end of the Rhaetian or in the earliest Hettangian were plants with ecological preferences that resulted in low preservation potential for their macroscopic remains, e.g. epiphytic plants or plants growing in sites with little potential for fast burial. From an ecological perspective, the crisis amongst land plants at the end of the Triassic was profound, with replacement or major disruptions of the terrestrial ecosystems (McGhee et al., 2013; Lindström, 2016). Recently, mass rarity, i.e. the reduction in the numerical abundances and/or the reduction in geographic ranges of several species, was suggested as a more robust measure of the severity of a biotic crisis than taxonomic extinction (Hull et al., 2015). In plant ecology and ecosystem analysis, rarity is a key factor when establishing whether a taxon is at risk of becoming endangered or extinct (Stohlgren and Kumar, 2013), but it is not often discussed for the big five biotic crises.

Here, two high resolution palynological records of the Danish Basin are used to assess whether mass rarity amongst land plants went extinct, the parent plant affinity is not fully resolved or even unknown (Lindström, 2016). Likely, many of the spore-producing plants that disappeared at the end of the Rhaetian or in the earliest Hettangian were plants with ecological preferences that resulted in low preservation potential for their macroscopic remains, e.g. epiphytic plants or plants growing in sites with little potential for fast burial. From an ecological perspective, the crisis amongst land plants at the end of the Triassic was profound, with replacement or major disruptions of the terrestrial ecosystems (McGhee et al., 2013; Lindström, 2016). Recently, mass rarity, i.e. the reduction in the numerical abundances and/or the reduction in geographic ranges of several species, was suggested as a more robust measure of the severity of a biotic crisis than taxonomic extinction (Hull et al., 2015). In plant ecology and ecosystem analysis, rarity is a key factor when establishing whether a taxon is at risk of becoming endangered or extinct (Stohlgren and Kumar, 2013), but it is not often discussed for the big five biotic crises.

Here, two high resolution palynological records of the Danish Basin are used to assess whether mass rarity amongst land plants played a role in the end-Triassic crisis scenario. The marine TJB succession of the cored Stenlille-1 and -4 wells on Sjælland, Denmark (Figure 1), are well constrained by palynology, δ²⁶⁰Corg-isotopes and other geochemical proxies, and can be correlated primarily with other NW European TJB strata, but also globally (Lindström et al., 2017b). Deposition took place in a shoreface to offshore setting in the northwestern part of an epicontinental sea that covered large parts of Europe. The two wells are situated 3.5 km apart on opposite sides of a gentle salt dome and exhibit minor variations in the sedimentary record. Two negative excursions in δ²⁶⁰Corg, the Marishi CIE and the Speleae CIE, can be correlated with TJB sections in the Tethys and Panthalassic oceans, where the last occurrence of typical Triassic and the first occurrence of typical Jurassic ammonoids bracket the marine mass extinction interval (Lindström et al., 2017b, 2021). Both these excursions have been interpreted to reflect massive volcanic degassing of 12C to the atmosphere from the CAMP, which could have acted as a trigger for the extinction (Hesselbo et al., 2002; Ruhl and Kürschner, 2011). In addition, Hg-anomalies in the succession provide a link to other Hg-records across the TJB and to the CAMP volcanism (Lindström et al., 2019; Lindström et al., 2021).

2 RESULTS

2.1 From Terrestrial Ecosystem Stability to Mass Rarity

The palynological records show that terrestrial vegetation of the extensive coastal areas and lowlands of the Danish Basin was remarkably stable during the middle Rhaetian (Figure 2). Variations in climate and sea level likely influenced vegetation changes on land as well as changes in depositional environment (Nielsen, 2003), contributing to abundance variations in the palynological records of the Gassum Formation (Figure 2). The pollen records show that cheirolepidiacean and taxodiaceous/cupressaceous conifers dominated the vegetation (Figure 2), occupying the upper canopy tier in drier and wetter environments, respectively (Tables 1 and 2; Figure 3A). Apart from these and Ricciisporites tuberculatus, most other typical Rhaetian tree pollen taxa were less common, but their parent plants were still important contributors to the upper canopy. The Rhaetian mire forests were most likely multi-storeyed and lush, with mid-canopy elements including very rare Erdtmanninthecales, rare Bennettitales and Ginkgoales/Cycadales, and common Caytoniales (Tables 1 and 2; Figure 3A), similar in composition to macroplant assemblages described from Greenland, southern Sweden and Germany (McElwain et al., 2007; Pott and McLoughlin, 2011; Van Konijnenburg-Van Cittert et al., 2018). Understorey plants, including ferns, lycophytes, bryophytes and sphenophytes, were generally rare to common and many of these plants were probably adapted to living in the shade of the trees or even as epiphytes on the trees (Figure 3A).

Two phases of mass rarity are evident in the Stenlille pollen records, here referred to as MR1 and MR2, respectively (Figure 2). MR1 occurs within the lowermost part of the Fjerritslev Formation (Figure 2) and encompasses the upper part of a black shale and the lowermost part of the succeeding “grey siltstone” interval of the Fjerritslev Formation (Lindström et al., 2017b) (Figure 2). At MR1, upper canopy plants appear to have been particularly hard hit. Both the coastal, salinity tolerant and xerophytic cheirolepidiaceous conifers, as well as the taxodiaceous/cupressaceous conifers that preferred wet habitats like mires and estuaries (Mussard et al., 1997), were severely
reduced in abundances within a c. 2 m thick interval from the middle of the black shale in the Stenlille-1 record (Figure 2A; Table 1). Four different cheirolepidiaceous pollen species exhibit rarity at this level commencing with *Granuloperculatipollis rudis*, followed in ascending order by *Classopollis meyerianus*, *Geopollis zwolinskai* and *C. classoides* (Figure 2). Rarity is also registered in *Rhaetipollis germanicus*, *Monosulcites minimus*, *Ovalipollis ovalis* and *Perinopollenites elatoides*, showing that other tree taxa were
TABLE 1 | Botanical affinity, habit, leaf morphology, pollination strategy, and environmental preferences for the parent plants affected at MR1 and MR2.

| Spore or pollen taxon | Mother plant Order Family | Habit | Leaf morphology | Pollen dispersal | Ecological preference | MR1 | MR2 |
|-----------------------|---------------------------|-------|-----------------|-----------------|-----------------------|-----|-----|
| Classopolis classoides | Conifer Coniferales Cheirolepidiaceae | Upper canopy | Scaly leaf shoots | Wind | Xerophytic, salinity tolerant | Non-permanent rarity | Recovery |
| Classopolis meyerianus | Conifer Coniferales Cheirolepidiaceae | Upper canopy | Scaly leaf shoots | Wind | Xerophytic, salinity tolerant | Non-permanent rarity | Recovery |
| Granuloperculatispolis rudis | Conifer Coniferales Cheirolepidiaceae | Upper canopy | Scaly leaf shoots | Wind | Xerophytic, salinity tolerant | Permanent rarity + Extirpation |
| Geopollis zwolinskai | Conifer Coniferales Cheirolepidiaceae | Upper canopy | Scaly leaf shoots | Wind | Xerophytic, salinity tolerant | Permanent rarity + Extirpation |
| Rhaetipollis germanicus | Gymnosperm Unknown | Unknown, here placed in mid-canopy | Unknown | Unknown | Unknown | Permanent rarity + Extirpation |
| Peninopollenites elatoides | Conifer Coniferales Taxodiaceae/ Cupressaceae | Upper canopy | Needle-like leaves | Wind | Moisture loving, mire | Non-permanent rarity | Recovery |
| Ovalipollis ovalis | Conifer Voltziales Unknown | Upper canopy | Needle-like leaves | Wind | Xerophytic | Permanent rarity | Extirpation |
| Monosulcites minimus | Seed fern Bennettitales Unknown | Mid canopy | Segmented or entire-margined | Insect | Deltaic and disturbed environments, mire | Non-permanent rarity | Recovery |
| Lunatisporites rhaeticus | Seed fern or Conifer Voltziales Unknown | Upper canopy | Needle-like leaves | Wind | ?Xerophytic | Permanent rarity | Permanent rarity + Extirpation |
| Semiretisporis gothae | Lycophyte Unknown | Ground cover | — | Water | Lowlands, hygrophytic | — | Permanent rarity + Extirpation |
| Cingulizonates rhaeticus | Lycophyte Unknown | Ground cover | — | Water | Lowlands, hygrophytic | — | Permanent rarity + Extirpation |
| Limbosporites lundbladiae | Lycophyte Unknown | Ground cover | — | Water | Lowlands, hygrophytic | — | Permanent rarity + Extirpation |
| Polyopodisporites polymicroforatus | Filicales Schizaceae | Ground cover | — | Water | Disturbed environments, ? hygrophytic | — | Permanent rarity + Extirpation |
| Ricciisporites tuberculatus | Gymnosperm Unknown, here placed in mid-canopy | Unknown | Unknown | Insect? Animal? | Unknown | — | Permanent rarity + Extirpation |

also affected (Figure 2A). The affinity of R. germanicus is unknown, but O. ovalis may have been produced by voltzialean conifers (Scheuring, 1970), and M. minimus by a bennettitalean parent plant (Pott et al., 2016). The Cheirolepidiaceae, taxodiacean/cupressacean and voltzialean conifers had scale- or needle-like leaves that would make them more fire prone than e.g., bennettitaleans that had elongated entire-margined or segmented leaves (Belcher et al., 2010) (Table 1). Wildfires under high temperatures indicative of crown fires in fire-prone vegetation are known to have occurred in the mid-Rhaetian mires (Petersen and Lindström, 2012). Palynofacies analyses of the Stenlille-4 samples show some intervals in the Gassum Formation where increased amounts of black phytoclasts and wood remains indicate that more intense wildfires may have occurred at times (Figure 2B). Rhaetian bennettitaleans seem to have had an ecological preference for highly disturbed and deltaic habitats (Pott, 2014), and M. minimus was found to be common to abundant in the fire struck Rhaetian mires of the Danish Basin (Petersen and Lindström, 2012). Except for the Bennettitaleae that may have been insect pollinated (Pott, 2014), the tree taxa affected at MR1 are all considered to have been wind pollinated (Table 1). In the Stenlille-4 record MR1 appears to be extended from the uppermost part of the Gassum Formation to lower part of the “grey siltstone” interval (Figure 2B). Discrepancies between the upper part of MR1 in this record and Stenlille-1 likely relates to the fact that the base of the “grey siltstone” in Stenlille-4 contains clay clasts probably derived from the black shale (Figure 2). Apart from the probable reworking of black shale material into the lowermost “grey siltstone”, the overall pattern is similar in the two wells. Although neither the parent plants of Classopolis classoides, C. meyerianus, Peninopollenites elatoides nor
**Monosulcites minimus** were extirpated in the end-Triassic extinction, the cheirolepidiacean conifers that produced *G. zwolinkai* and *G. rudis*, as well as the voltzialean conifer that shed *O. ovalis* and the unknown parent plant of *R. germanicus*, were all victims of the crisis.

### 2.2 Ecological Upheaval and Pioneering Plants

The ecological disruption during MR1 resulted in a major reduction of the upper and mid-canopy tree elements, suggesting major deforestation in the area and a marked shift towards a more open landscape (*Tables 1* and *2; Figure 3B*). Marked and consistently increased levels of black phytoclasts and black wood remains, and significantly decreased amounts of spores and pollen compared to prior to MR1 may indicate that increased wildfire (*Figure 2B*). Both the vegetation density and volume were likely severely decreased during this floral crisis interval as indicated by low TOC values (Lindström et al., 2019), and by significantly lower amounts of spores and pollen during this interval (*Figure 2B*). The deforestation allowed some previously less prominent taxa to flourish during the deposition of the “grey siltstone” (*Figures 1, 2*). In particular the understorey spore-producing schizaeacean (*Polypodiisporites polymicroforatus* spores) and dipterid and/or matoniaceous ferns (*Van Konijnenburg-Van Cittert et al., 2020*) (*Deltoidospora* spores; not shown on Figure 2) seem to have benefited from the open landscape (*Table 2; Figure 3B*). These opportunistic ferns had a significantly faster life cycle than trees and could perhaps reproduce both sexually and through asexual spreading via rhizomes. These ferns flourished together with an unknown, possibly ruderal, gymnosperm (*Ricciisporites*...
tuberculatus pollen) (Kürschner et al., 2014) that also increased in abundance compared to in the Gassum Formation. A similar vegetation response has been documented across Europe and interpreted as reflecting widespread distribution of pioneering plant communities after deforestation (van de Schootbrugge et al., 2009; Lindström, 2016; Lindström et al., 2017b; Gravendyck et al., 2020). Interestingly, Monosulcites minimus is significantly decreased during the deforestation interval, i.e. the “grey siltstone” beds, despite the parent plants possible preference to disturbed habitats (Pott, 2014) (Figure 2). Some typical Rhaetian spore-pollen taxa, e.g., the lycophyte spores Limbosporites lundbladiae and Cingulizonates rhaeticus, and the probable conifer or seed fern pollen Lunatisporites rhaeticus, show none or only marginal changes in abundance during this interval (Figure 2). Another typical Rhaetian lycophyte spore, Semiretisporis gothae, even increased in abundance during the deposition of the “grey siltstone” beds.

2.3 Mass Rarity and Recovery
In the uppermost part of the “grey siltstone” beds, a second mass rarity event, MR2, was instigated with the virtual disappearance of the low-abundance species S. gothae, L. rhaeticus, C. rhaeticus, and L. lundbladiae (Figures 2, 3C). Two of the opportunistic palynofloral elements in the “grey siltstone” beds, P. polymicroforatus and R. tuberculatus, also show marked abundance losses, reflecting that the collapse of the pioneering ecosystem occurred during MR2 (Figure 2). MR2 coincided with the onset and first negative peak of the Spelae CIE (Figure 2), which coincides with a sudden relative increase in the equisetalean spore Calamospora tener and in the freshwater algae Botryococcus braunii (CCM zone in (Lindström et al., 2017b)). As equisetaleans prefer wet environments like riverbanks or lake shores, this signals an increase in freshwater input at that time. After this there was a return of, in particular, the taxodiacean/cupressacean (P. elatoides pollen), but also to a lesser extent cheirolepidiaceous (C. meyerianus and C. classoides pollen) and bennettitalean (M. minimus pollen) trees (Figure 2). Although cupressacean conifers appear to have returned in large numbers signalling re-establishment of coastal mires, the drier coastal habitats that supported the cheirolepidiaceous conifers appear to have been greatly diminished (Tables 1 and 2; Figure 3C). In general, the post-crisis vegetation appears less dense in the upper canopy while the mid-canopy tier seems almost eradicated, similar to macroplant fossil observations in East Greenland (McElwain et al., 2007). The post-crisis vegetation contained several rare or ghost taxa, the latter including e.g., G. rudis, O. ovalis, L. rhaeticus, P. polymicroforatus and R. tuberculatus, that were victims of the crisis (Figure 3C). Later, Pinuspollenites minimus, a previously rare pineaceous conifer pollen increased in abundance during the post-crisis and recovery interval, probably inhabiting hinterland or drier areas (Lindström, 2016) (Figure 3C).

2.4 Diversity Patterns Reveal Onset of Disturbances
Diversity indices provide additional clues to the timing and severity of the environmental changes that led to the mass rarity phases. In accordance with the intermediate disturbance hypothesis (IDH) (Svensson et al., 2012), two intervals where species richness increased and peaked indicate that the ecosystem was subjected to disturbances of intermediate magnitude and/or frequency; namely intermediate disturbance 1 (ID1) and 2 (ID2) (Figure 4). At ID1 species richness increased and peaked just prior to and during the lower part of MR1 (Figure 4), indicating an intermediate level of disturbance that hindered competitive exclusion by suppressing species that were previously dominant and allowing new species to colonize free space (Svensson et al., 2012). This is reflected by the initial increase in richness coinciding with a minimum in dominance, <0.1, and an increase in evenness to >0.8 (Figure 4). Just prior to MR1, while species richness was still high, evenness dropped continuously to the top of MR1 (Figure 4), indicating high level disturbances promoting dominance of a few disturbance specialists, namely P. polymicroforatus, R. tuberculatus and Deltoidospora spp. (Figure 2). The decline in evenness continued to the top of MR1 as species richness also declines, reflecting a subsequent loss in biodiversity (Figure 4).

At ID2 in the upper part of the “grey siltstone” beds, richness again reached maximum levels indicating intermediate disturbance (Figure 4). At the top of the beds, richness fell successively and reached a minimum during MR2 (Figure 4). This coincided with a fall in dominance and a rise in evenness, suggesting that the taxa that were ecologically dominant during the “grey siltstone” beds were now suffering, whereas other species were recovering or establishing. The diversity indices suggest that conditions improved after the Spelae CIE and the palynology suggests that some of the previously lost coastal habitats were restored, as primarily taxodiacean/cupressacean, but also to lesser extent cheirolepidiaceous, conifers increase in abundance (Figures 1–3).

3 DISCUSSION
3.1 Comparison With Other Records
3.1.1 St. Audrie’s Bay, United Kingdom
The St. Audrie’s Bay succession (Figures 1, 5) is one of the extensively studied key localities for the TJB, and it is well constrained by palynology, ammonoids and organic C-isotopes (Hesselbo et al., 2002; Hesselbo et al., 2004). Hesselbo et al. (2004) indicated that a maximum flooding surface (MFS) is present in the middle of the Westbury formation, around the level of the Marshi CIE, but they could not specify the exact location. Maximum abundances of dinoflagellate cysts (Bonis et al., 2010) indicate that the MFS could be located higher, in the uppermost part of the Westbury Formation and lower to upper Cotham Member of the Lilstock Formation. It is within this interval that several spore-pollen taxa begin to decline markedly in abundance. Granuloperculatipollis rudis is the first taxon to exhibit rarity already in the uppermost Westbury Formation. Then Perinopollenites elatoides, which was never common, declines in abundance and Ovalipollis ovalis becomes rare in the uppermost lower Cotham Member and at this level Classopollis meyerianus also shows a marked decline in abundance, although it never becomes rare (Bonis et al., 2010,
FIGURE 4 | Comparison of species richness, dominance (1-Simpson index), and Pielou’s evenness in Stenlille-1 and Stenlille-4. ID1 and ID 2 refer to two intervals with increased to maximum values in species richness indicate intermediate disturbance (Connell, 1978; Svensson et al., 2012). Hg and Hg/TOC records have been interpreted to reflect volcanic pulses of the CAMP (Lindström et al., 2021). Thin stippled red lines indicate mean level of the respective diversity indices.

FIGURE 5 | Correlation of δ^{13}Corg-records, mass rarity phases (MR1 and 2) and crisis interval [i.e. Polypodiisporites polymicroforatus abundance interval of Lindström et al. (2017b)] in Stenlille-1, St. Audrie’s Bay, Bonenburg, Maniental and Kuhjoch. The recorded taxa are also shown in Table 3. Numbers after the localities refer to the following references: 1. Lindström et al. (2017b), 2. Lindström et al. (2021), 3. Hesselbo et al. (2002), 4. Hesselbo et al. (2004), 5. Bonis et al. (2010), 6. Graven et al. (2020); Heusnisch et al. (2010), 8. Bonis et al. (2009), 9. Hillebrandt et al. (2013). The correlation is based on Lindström et al. (2017b), Lindström et al. (2021).
TABLE 2 | Mean percentage data for the various spore-pollen taxa or groups of taxa shown in Figure 3 as well as assumed forest level and climate preference.

| Nr. on Figure 3 | Taxon or group of taxa | Pre-extinction | Crisis | Recovery and post-extinction | Forest level | Climate preference |
|-----------------|------------------------|----------------|--------|-----------------------------|--------------|-------------------|
| 1               | Pollen Classopollis classoides | 12.4% | 0.2% | 1.3% | Upper canopy | Xerophytic |
| 2               | Pollen Classopollis meyerianus | 3.9% | 0.3% | 1.8% | Upper canopy | Xerophytic |
| 3               | Pollen Geopollis zwolinskai | 1.3% | 0.0% | 0.0% | Upper canopy | Xerophytic |
| 4               | Pollen Granuloperculatipollis rudis | 5.5% | 0.0% | 0.1% | Upper canopy | Xerophytic |
| 5               | Pollen Lunatisporites raeticus | 1.0% | 0.3% | 0.1% | Upper canopy | Xerophytic |
| 6               | Pollen Monosulcites minimus | 1.8% | 0.2% | 0.3% | Mid canopy | Wet |
| 7               | Pollen Ovalipollis ovalis | 1.2% | 0.1% | 0.1% | Upper canopy | Xerophytic |
| 8               | Pollen Perinopollenites elatoides | 18.8% | 4.2% | 27.4% | Upper canopy | Wet |
| 9               | Pollen Rhaetipollis germanicus | 0.9% | 0.1% | 0.0% | Unknown | Unknown |
| 10              | Pollen R. tuberculatus | 14.1% | 16.9% | 0.7% | Unknown | Xerophytic |
| 11              | Spores Polypodisporites polymicroforatus | 1.0% | 15.9% | 0.6% | Ground cover | Drier |
| 12              | Spores Cingulizonates raeticus | 0.3% | 0.6% | 0.1% | Ground cover | Wet |
| 13              | Spores Limbosporites lundbladiae | 0.3% | 0.4% | 0.1% | Ground cover | Wet |
| 14              | Spores Semiretisporis gothae | 0.0% | 0.4% | 0.0% | Ground cover | Wet |
| 15              | Pollen Pinuspollenites minimus | 0.7% | 0.4% | 8.1% | Upper canopy | Xerophytic |
| 16              | Pollen Vitreospores spp. | 3.3% | 1.3% | 0.9% | Mid canopy | Xerophytic |
| 17              | Pollen Other bisaccates | 8.0% | 5.7% | 2.9% | Upper canopy | Xerophytic |
| 18              | Pollen Alisporites spp. | 2.3% | 1.8% | 1.1% | Upper canopy | Xerophytic |
| 19              | Pollen Other non-saccate conifers | 1.7% | 0.6% | 2.0% | Upper canopy | Xerophytic |
| 20              | Pollen Monosulcites spp. other | 1.3% | 0.7% | 1.9% | Upper canopy | Unknown |
| 21              | Pollen Eucommiidites spp. | 0.4% | 0.0% | 0.1% | Mid canopy | Drier |
| 22              | Spores Deltoisporites spp. | 8.7% | 26.5% | 23.8% | Ground cover | Drier |
| 23              | Spores Other ferns (minus Osmundaceae) | 3.9% | 8.3% | 7.2% | Ground cover | Wet |
| 24              | Spores Other lycophytes | 1.7% | 2.2% | 1.9% | Ground cover | Wet |
| 25              | Spores Calamospora spp. | 0.5% | 1.9% | 5.0% | Ground cover | Wet |
| 26              | Spores Bryophytes total | 1.8% | 2.2% | 4.9% | Ground cover | Wet |
| 27              | Pollen Chasmatosporites spp. | 1.2% | 1.7% | 1.2% | Upper canopy | Drier |
| 28              | Spores Osmundaceae total | 1.3% | 3.9% | 6.0% | Ground cover | Wet |

Figures 3, 4 therein). *Rhaetipollis germanicus* becomes rare just at the level marked by desiccation cracks (see below) and this is also where *C. classoides* first drops markedly in abundance (Bonis et al., 2010) (Figure 5). The presence of an erosion surface with prominent desiccation cracks up to 90 cm deep at the top of the lower Cotham member indicates a hiatus at that level in the succession (Hesselbo et al., 2004) (Figure 5). Sedimentary structures including flat-topped ripples indicate water depth of just a few meters in the upper Cotham Member of the Lilstock Formation and thus the culmination of a regressive event (Hesselbo et al., 2004). This corresponds stratigraphically to the sea-level fall in the Danish and North German Basins. Thus, a first phase of mass rarity (MR1) can be recognized in the lower Cotham Member, below the desiccation cracks (Table 3; Figure 5). It is possible that most of the crisis interval at St. Audrie’s Bay is missing (Lindström et al., 2017b). In addition, several of the taxa reviewed herein where not listed in the palynological record from St. Audrie’s Bay, namely *Cingulizonates raeticus*, *Limbosporites lundbladiae*, *Semiretisporis gothae*, and *Geopollis zwolinskai* (Bonis et al., 2010) (Table 3), so their responses cannot be assessed. A second mass rarity (MR2) can be distinguished within a narrow interval above the desiccation cracks, encompassing the onset of the Spelae CIE, with rarity in *Lunatisporites raeticus*, *Perinopollenites elatoideae*, *Polypodisporites polymicroforatus* and *Riccisporites tuberculatus* (Table 3; Figure 5). The MR2 interval in St. Audrie’s Bay corresponds more or less to the interval referred to as the “lower CIE” in by Fox et al. (2020), interpreted to represent a dramatic perturbation to the ecosystem by shallowing with increasing freshwater conditions and development of microbial mats. The onset of the Spelae CIE is in St. Audrie’s Bay also associated with an increased abundance of the equisetalean spore *Calamospora tener*, which may suggest increased runoff from a nearby freshwater source where equisetalean plants may have thrived along riverbanks or lake shores.

3.1.2 Bonenburg, Germany
In the Bonenburg section (Figures 1, 5), located in the German Basin, six taxa exhibit mass rarity during an interval stratigraphically equivalent with MR1, namely *Classopollis classoides*, *C. meyerianus*, *Geopollis zwolinskai*, *Granuloperculatipollis rudis*, *Ovalipollis ovalis* and *Rhaetipollis germanicus* (Gravendyck et al., 2020) (Table 3; Figure 5). The abundance drops in these taxa occur at or immediately after a peak in marine phytoplankton abundance, in particular the fully marine dinoflagellate cyst taxon *Rhaetogonyaulax raeticia* in the upper part of the Contorta Beds, which suggests that this level represents a maximum transgression (Gravendyck et al., 2020). Only *G. rudis* disappears at this level. *Perinopollenites elatoideae* is rare throughout the section and appears to have been unaffected. *Lunatisporites raeticus* is also rare but drops in abundance within the crisis interval (Triletes beds) (Gravendyck et al., 2020) (Table 3; Figure 5). At a level equivalent to MR2,
TABLE 3 | Comparison of mass rarities, extinctions and recoveries of the selected taxa on Figure 1, for Stenlille-1 and -4 (this paper), St. Audrie's Bay (United Kingdom; Bonis et al., 2010), Bonenburg (Germany; Gravendyck et al., 2020), Mariental (Germany; Heunisch et al., 2010), Kuhjoch (Austria; Bonis et al., 2009; Hillebrandt et al., 2013). The localities are correlated according to ref. 1010. In the case of Mariental, only Classopollis spp. is listed. Numbers for the seven lowest lines in the table refer to number of taxa affected during each phase.

| Selected taxa | Stenlille-1 | Stenlille-4 | St. Audrie’s bay | Bonenburg | Mariental | Kuhjoch |
|--------------|-------------|-------------|-----------------|-----------|-----------|--------|
|              | MR1 | Crisis | MR2 | Post-crisis | MR1 | Crisis | MR2 | Post-crisis | MR1 | Crisis | MR2 | Post-crisis | MR1 | Crisis | MR2 | Post-crisis | MR1 | Crisis | MR2 | Post-crisis |
| Classopollis classoides | x | — | — | R | x | — | — | R | (x) | — | — | R | x | — | R | — | — | (x) | R | — |
| Classopollis meyerianus | x | — | — | R | x | — | — | R | (x) | — | — | R | x | — | R | — | x | — | — | — |
| Geopollis semiannulata | x | — | — | — | x,† | — | — | — | x,† | — | not listed | — | — | — | x,† | — | — | — | — |
| Granuloperculatipollis rudis | x,† | — | — | — | x,† | — | — | — | x,† | — | not listed | — | — | — | x,† | — | — | — | — |
| Rhaetipollis germanicus | x,† | — | — | — | x,† | — | — | — | x,† | — | not listed | — | — | — | x,† | — | — | — | — |
| Monosulcites minimus | x | — | — | R | x | — | R | — | — | x,† | — | — | R | — | — | x,† | — | — | — | — |
| Ovalipollis ovalis | x | — | — | — | x,† | — | — | — | x,† | — | not listed | — | — | — | x,† | — | — | — | — |
| Reticulopollis tuberculatus | x | — | — | — | x,† | — | — | — | x,† | — | not listed | — | — | — | x,† | — | — | — | — |
| Cingulizonates rhaeticus | x | — | — | — | x,† | — | — | — | x,† | — | not listed | — | — | — | x,† | — | — | — | — |
| Semiretisporis gothae | x | — | — | — | x,† | — | — | — | x,† | — | not listed | — | — | — | x,† | — | — | — | — |
| Polypodiisporites polymicroforatus | x | — | — | — | x,† | — | — | — | x,† | — | not listed | — | — | — | x,† | — | — | — | — |

R = recovery.

x = rarity.

† = major decline.

† = extinction.
Limbosporiteslundbladiae, Polypodiisporitespolymicroforatus, Semiretisporis gothae and Riccisporites tuberculatus exhibit mass rarity, while Classopollisclassoides, C. meyerianus and G. zwolinskai again increase in abundance (Table 3; Figure 5). The recovery of G. zwolinskai is surprising, but this has also been documented in Mariental (see below). This contrasts to records further north where this taxon appears to have gone extinct during MR1 (Lund, 2003; Lindström et al., 2017a; this paper). Semiretisporis gothae and R. germanicus went extinct at this level. Ovalipollisovalis, Lunatisporitesrhaeticus and Limbosporiteslundbladiae lingered on but disappeared during the post-crisis interval (Table 2; Figure 5). Riccisporites tuberculatus increased again towards the top of the record (Gravendyck et al., 2020) similar to what is seen in Stenlille (Figure 2), and most likely went extinct at a higher level.

### 3.1.3 Mariental, Germany

In the Mariental succession (Figure 1), six taxa show marked abundance drops at a level equivalent to MR1 in the upper part of the Contorta Beds, namely Classopollisspp., G. zwolinskai, Limbosporiteslundbladiae, Ovalipollisovalis, Perinopolleniteselatoides and Rhaetipollis germanicus (Heunisch et al., 2010). The latter as well as Granuloperculatipollisrudis went extinct at that level (Figure 5). A similar pattern was also noted by Lund (2003) in the German well Eitzendorf 8 located in the Bremen area, where a MFS in the upper part of the Mittel-Rhät Contorta Beds is marked by an acme of R. rhaetica, and succeeded by the last occurrences of G. rudis, G. zwolinskai and R. germanicus (Lund, 2003). During the middle part of the crisis interval (the Triletes beds) Riccisporites tuberculatus and Lunatisporitesrhaeticus exhibited mass rarity (Heunisch et al., 2010) (Table 3; Figure 5). The latter as well as Limbosporiteslundbladiaedisappeared at a level equivalent to MR2, where also Polypodiisporitespolymicroforatus and Semiretisporisgothae show mass rarity. Four taxa, O. ovalis, P. polymicroforatus, R. tuberculatus and Semiretisporisgothae went extinct during the post-crisis interval (Heunisch et al., 2010) (Table 2; Figure 5). Classopollisspp., Perinopolleniteselatoides and Geopolliszwalinskai all recovered during MR2 (Heunisch et al., 2010).

### 3.1.4 Kuhjoch, Austria

The Triassic–Jurassic boundary succession at Kuhjoch (Figure 1 and) is the Global Stratigraphic Section and Point (GSSP) for the base of the Jurassic (Hillebrandt et al., 2013). Several of the herein compared taxa were not listed in the Kuhjoch record, including G. zwolinskai, G. rudis, M. minimus and S. gothae (Table 3; Figure 5), and two taxa, P. elatoides and R. tuberculatus were seemingly unaffected (Hillebrandt et al., 2013). The open marine Kössen Formation in succeeded by a sea-level fall and the deposition of the grey Tiefengraben Member of the Kendlback Formation (Hillebrandt et al., 2013). Major drops in abundances occur within an interval corresponding to MR1 affecting e.g. C. meyerianus, C. classoides, while G. rudis and R. germanicus exhibit mass rarity (Bonis et al., 2009) (Table 3; Figure 5). During the subsequent crisis interval, i.e. the Schattwald beds deposited during continued sea-level fall, mass rarity is registered in L.lundbladiae, O. ovalis and L. rhaeticus. There instead, Polypodiisporitespolymicroforatus and other spores dominated the assemblages (Bonis et al., 2009). Granuloperculatipollisrudis and R. germanicus went extinct during MR1, while Lunatisporitesrhaeticus was extirpated during the crisis interval (Table 3; Figure 5). At a narrow interval at the base of the Spelae CIE, Polypodiisporitespolymicroforatus, Limbosporiteslundbladiae, Lunatisporitesrhaeticus and Semiretisporisgothae became rare and later disappeared during the earliest Jurassic (Table 3; Figure 5).

### 3.1.5 A Supraregional Pattern

The comparison with other high-resolution palynological records from the European epicontinental sea and the northern Tethys margin clearly shows that mass rarity in coastal and near-coastal lowland plants occurred in two phases on a supraregional scale. Although significant differences in the pollen records occur for some species, e.g. for P. elatoides, the taxodiacean/cupressacean plant parent of which is known to have survived the end-Triassic crisis. However, while it was severely suppressed during MR1 and in both the “grey siltstone” beds in the Stenlille wells and the Triletes beds of the German well Mariental (Heunisch et al., 2010), the parent plant of P. elatoides occurred consistently in low numbers and appeared unaffected at the other localities (Table 3). This is most likely related to the ecological and climatological preference of the taxodiacean/cupressacean parent plant, which was never abundant during the Rhaetian in the United Kingdom or Austria (Bonis et al., 2010) (Hillebrandt et al., 2013), as well as the depositional environment at each site.

### 3.2 The Onset of an Ecological Crisis

The magmatic activity in the Central Atlantic Magmatic Province may have affected the terrestrial ecosystem in multiple ways, following the flow chart in Figure 6. The successive abundance losses at MR1 were unprecedented during the Rhaetian of NW Europe. The preceding increase in richness, ID1, which signals intermediate levels of ecosystem stress, coincided with and immediately succeeded the Marshi CIE, believed to reflect added excess of light carbon to the carbon cycle as greenhouse gases (Hesselbo et al., 2002; Ruhl and Kürschner, 2011; Lindström et al., 2021), which may have included methane from magma-sediment interaction (Ruhl et al., 2011; Capriolo et al., 2021) (Figure 4). The local/regional effects of the subsequent global warming would likely have varied depending on the prevailing palaeoclimatic conditions at each site. The maximum values in species richness also coincided with increased Hg-loading and moderately increased occurrences of mutated fern spores (Lindström et al., 2019) interpreted to reflect volcanic emissions from CAMP volcanism (Thibodeau et al., 2016; Percival et al., 2017; Lindström et al., 2019; Lindström et al., 2021). Increased occurrences of mutatedClassopolisspores at a corresponding stratigraphical level in Germany (Gravendyck et al., 2020) further suggests that mutagenic stress also affected the cheirolepidiacean conifers. The mutagenesis could have been caused by Hg-toxicity as suggested for the ferns (Lindström et al., 2019), however, other heavy metal toxicities or increased UV-B
radiation, as suggested for plant mutagenesis at the end-Permian event (Benca et al., 2018; Chu et al., 2021), cannot be excluded at this time (Figure 6).

3.3 The Effects of Rapid Sea-Level Changes

The coastal and near-coastal lowland terrestrial ecosystems along the margins of the European continental sea and the northern Tethys margin were vulnerable to sea-level changes, but the vegetation appears to have remained relatively stable during the middle Rhaetian. The black shale at the base of the Fjerritslev Formation was deposited during the late Rhaetian transgression (marked by a maximum flooding surface, MFS7; Figure 2) (Nielsen, 2003; Lindström and Erlström, 2006) which in contrast to previous transgressive events was associated with massive abundance losses in trees growing in coastal habitats or near-coastal lowlands during MR1 (Figure 2). The fact that this mass rarity occurred just prior to, during and after the culmination of this late Rhaetian transgression indicates that the rate of sea level rise outpaced the organic and inorganic sediment supply in the coastal region, leading to habitat fragmentation of the terrestrial coastal and lowland ecosystems to such an extent that it instigated the mass rarity (Figure 6). The fast relative sea-level rise in combination with global warming (Ruhl et al., 2011), increased wildfire frequency (Petersen and Lindström, 2012) and heavy metal pollution (Lindström et al., 2019) likely exerted profound stress on the terrestrial ecosystem at that time (Figure 6).

The late Rhaetian transgression was succeeded by a prominent regression during which the "grey siltstone beds" were deposited (Figure 2). This major drop in sea-level in the late Rhaetian was part of a regression-transgression couplet that can be recognized over large parts of Europe (Hallam, 1997; Hallam and Wignall, 1999), but the mechanism behind this sea-level drop remains to be explained. Hallam (1997) favoured a tectono-eustatic
mechanism other than sea-floor spreading, comparing it in pace and magnitude to glacio-eustacy. Hallam and Wignall (1999) suggested that the regression-transgression couplet was linked to the formation of the CAMP, and widespread occurrences of seismites in co-eval strata across Europe (Simms, 2003; Lindström et al., 2015) seem to support this. In the Danish Basin, this regression culminated with coastal deposition far into the basin, and in some marginal areas fluvial erosion and incision took place (Nielsen, 2003). As a consequence of this, late Rhaetian, coincident with increased wild fires, decreased tree density (Petersen and Lindström, 2012), and pyrolytic polycyclic aromatic hydrocarbons (PAHs) (van de Schootbrugge et al., 2009; Lindström et al., 2021). In addition, PAH formed by incomplete combustion of organic matter suggest that intrusive coking of organic-rich sediments in the CAMP area also occurred (van de Schootbrugge et al., 2009). At the same level (Lindström et al., 2021), cuticles from bennettitaleans and ginkgoaleans in Greenland exhibit lesions and distortions typical of sulfuric acid deposition and this was further intensified across the TJB (Steinthorsdottir et al., 2018), suggesting ongoing sulphur-rich volcanic emissions from the CAMP, in line with that some CAMP lavas contained especially high levels sulphur compared to other LIPs (Callegaro et al., 2014). Acid rain would have exacerbated the physiological responses that TJB plants had to ongoing heat and water stress, possibly further increasing the hydrological cycle (Steinthorsdottir et al., 2012), which could explain why the opportunistic plants adapted to drought began to decline in abundance. MR2 preceded the prominent Spelae CIE (Figures 2, 4). This negative excursion has been interpreted to reflect atmospheric increase in light carbon either through degassing of volcanic CO$_2$ or methane (Hesselbo et al., 2002; Ruhl et al., 2011). However, this was recently questioned by Fox et al. (2020) who instead suggested that the negative CIEs at the Spelae level were regional phenomena caused by fractionation by microbial mats under brackish water conditions. An increase in freshwater algae in the Stenlille record also seems to indicate enhanced runoff (Lindström et al., 2012), but because negative CIEs at the Spelae level are also present outside the European epicontinental sea, e.g. in Argentina (Ruhl et al., 2020), further studies are needed to resolve this issue. Regardless of which, the Spelae CIE did not instigate the MR2 mass rarity in land plants. Instead, living conditions for some of the plants that had suffered during MR1 (e.g. Classopolis, Perinopollenites elatoides) appear to have improved during the Spelae CIE, perhaps due to stabilizing climatic conditions. A second relative sea-level rise may have helped to restore some of the coastal and near-coastal lowland habitats that were lost during and after MR1 (Figures 2–5). Thus, at least in the northern hemisphere both the destruction and resurrection of coastal and near-coastal terrestrial ecosystems were tightly linked to relative sea-level changes.

3.4 Rare Species Are Especially Vulnerable to Climate and Environmental Change

Today, areas with long term stable climatic conditions have been found to contain higher numbers of rare plants, because of the reduced extinction risks (Enquist et al., 2019). However, because rare plants species are more susceptible to reductions in populations size, they are more likely to go extinct by orders of magnitude than more abundant species during times of rapid climate change (Enquist et al., 2019). The geographical and numerical reduction of plant species during the two end-Triassic mass rarity phases attest to the severe impact of this crisis on the vegetation in NW Europe.
The causality behind the mass rarity amongst the plants was clearly complex, including both direct effects of the volcanic activity in the CAMP, as well as subsequent feedback mechanisms. The volcanic activity of the CAMP emitted multiple stressors, including greenhouse gases, SO₂, Hg-pollution, PAHs, and likely also halocarbons (Figure 6). The results emphasize that rapid sea-level changes, probably driven by tectono-eustatic activity in the CAMP area, severely affected the coastal and near-coastal lowland vegetation, to the extent that many species already stressed by climate change and volcanic pollution, may have had difficulties coping with fragmentation of habitats and changes groundwater levels. The mass rarity was profound as it affected both already rare as well as previously abundant species and must have had a cascading negative impact on the terrestrial ecosystem as a whole, altering the conditions for a range of organisms from the lowest to the highest trophic levels. The mass rarity in plants was supraregional across NW Europe (Table 2), although small populations of typical Rhaetian plant taxa lingered without recovery into the earliest Jurassic (early Hettangian) in some areas. Most of these ghost taxa are only known by their spores or pollen and have not been identified as macroplant fossils, thus they were rare already prior to the crisis. Their extirpation in the aftermath of the end-Triassic crisis attest to the vulnerability of plant species to reductions in abundance and geographic range, especially in coastal regions. This should resonate when considering present day anthropogenic ecosystem disturbances and ongoing and future climate change scenarios (IPCC, 2021) that, similar to the end-Triassic crisis include rapid sea-level rise (De Conto et al., 2021) in combination with increased frequency and intensity of climate extremes such as prolonged heatwaves and drought, wildfires, major storms and heavy precipitation.

4 METHODS

4.1 Sampling, Palynological Processing and Analysis

109 sedimentary rock samples from Stenlille-1 and 87 ones from Stenlille-4 were processed using standard palynological methods at the Palynology Laboratory at GEUS. Approximately 20 g of each sample was crushed and treated with hydrochloric and hydrofluoric acids to remove carbonate and silicate minerals, respectively. Heavy minerals were removed from the residues using heavy liquid separation. The organic residues were mildly oxidized using nitric acid and filtered on 11 μm mesh filters. Strew slides were prepared after each step in the preparation after the hydrofluoric acid treatment, using glycerine gel as mounting medium. The palynomorph content of the samples were assessed by counting 300 specimens in one representative strew slide from each sample. Spore and pollen taxa abundances were normalized against the total spore-pollen count in each sample. It is important to remember that spore/pollen production in a plant is not necessarily related to the actual abundance of that plant in the ecosystem, and the parent plant of continuously rare spore/pollen taxa may still have played an important role in the ecosystem. In addition, plants that rely on insect or animal assistance with dispersal and pollination may have a more restrained pollen production compared to those that rely on wind for dispersal. In the late Triassic, prior to the diversification of the angiosperms, possible insect pollination has only been suggested for the Bennettitales (Labandeira et al., 2007).

Seventy-five samples from Stenlille-4 were analysed for palynofacies. Two-hundred particles or more were counted from each sample on the slide prepared after heavy liquid separation but prior to swirling. The palynodebris and palynomorphs were divided into ten different categories: Black opaque phytoclasts which include opaque usually rounded particles; black wood encompasses all black wood remains, both blade shaped, needle shaped and equidimensional ones; plant tissue includes all non-woody and non-cuticular plant remains; cuticles include translucent to semi-translucent tissue with cellular structures with or without stomata; Amorphous organic matter (AOM) encompasses degraded material either as structureless amorphous or structured amorphous matter; marine phytoplankton includes acritarchs, dinoflagellate cysts and prasinophytes, however, for graphic purposes unidentifiable palynomorphs were also included in this group but are listed separately in Supplementary Table S3; freshwater microalgae primarily include Botryococcus braunii and zygnematacean cysts; Spores also include rare fragments of clitellate cocoon fragments; and finally pollen. The counts are listed in Supplementary Table S3.

4.2 Mass Rarity Assessment

In palynology, loss of species abundance has often been used as an indicator of biotic crises without specifically referring this to mass rarity (Twitchett et al., 2001; Lindström and McLoughlin, 2007; van de Schootbrugge et al., 2009). The mass rarity of a taxon was determined as either: 1) a marked shift from consistently common to abundant abundance values to consistently or inconsistently rare occurrences, i.e. the last common occurrence (LCO), or 2) as a marked shift from consistently present in low abundances to inconsistently present in low or lower abundances, i.e. the last consistent occurrence (LCon). In fact, mass rarity is routinely used in palynostratigraphy as the last common occurrence (LCO) of a taxon, or for taxa of low abundance the last consistent occurrence (LCon) may also reflect mass rarity in the parent plant.

4.3 Diversity Patterns

Diversity trends in the successions were analyzed using species richness, dominance (1-Simpson index) and Pielou’s evenness calculated using the software PAST4.05. Because many of the spore- and pollen-genera at the Triassic–Jurassic boundary are monospecific, species richness is only marginally higher than genus richness. Reworked taxa were excluded from the diversity analyses. The diversity patterns were assessed primarily according to the intermediate disturbance hypothesis (Connell, 1978; Svensson et al., 2012).
DATA AVAILABILITY STATEMENT
The original contributions presented in the study are included in the article/Supplementary Material, further inquiries can be directed to the corresponding author.

AUTHOR CONTRIBUTIONS
SL designed the project, produced and analysed the data, and wrote the manuscript.

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