Field rates of Sivanto™ (flupyradifurone) and Transform® (sulfoxaflor) increase oxidative stress and induce apoptosis in honey bees (Apis mellifera L.)

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Abstract

Pesticide exposures can have detrimental impacts on bee pollinators, ranging from immediate mortality to sub-lethal impacts. Flupyradifurone is the active ingredient in Sivanto™ and sulfoxaflor is the active ingredient in Transform®. They are both relatively new insecticides developed with an intent to reduce negative effects on bees, when applied to bee-attractive crops. With the growing concern regarding pollinator health and pollinator declines, it is important to have a better understanding of any potential negative impacts, especially sub-lethal, of these pesticides on bees. This study reports novel findings regarding physiological stress experienced by bees exposed to field application rates of these two insecticides via a Potter Tower sprayer. Two contact exposure experiments were conducted—a shorter 6-hour study and a longer 10-day study. Honey bee mortality, sugar syrup and water consumption, and physiological responses (oxidative stress and apoptotic protein assays) were assessed in bees exposed to Sivanto™ and Transform®, and compared to bees in control group. For the longer, 10-day contact exposure experiment, only the Sivanto™ group was compared to the control group, as high mortality recorded in the sulfoxaflor treatment group during the shorter contact exposure experiment, made the latter group unfeasible to test in the longer 10-days experiment. In both the studies, sugar syrup and water consumptions were significantly different between treatment groups and controls. The highest mortality was observed in Transform® exposed bees, followed by the Sivanto™ exposed bees. Estimates of reactive oxygen/nitrogen species indicated significantly elevated oxidative stress in both pesticide treatment groups, when compared to controls. Caspase-3 protein assays, an indicator of onset of apoptosis, was also significantly higher in the pesticide treatment groups. These differences were largely driven by post exposure duration, indicating sub-lethal impacts. Further, our findings also emphasize the need to revisit contact exposure impacts of Sivanto™, given the sub-lethal impacts and mortality observed in our long-term (10-day) contact exposure experiment.
Introduction

Pesticides can negatively impact managed bee stocks and native bee populations [1–8]. These impacts include lethal effects, as well as a host of sub-lethal effects, including cognitive impairments [6,7,9–13], flight and homing disruptions [3, 14,15] and detrimental physiological alterations [16–19]. Although the underlying physiological mechanisms that cause direct lethality of insecticides to bees are well-resolved, the connection between physiological processes and sub-lethal effects are not well understood. The lethal toxicity is currently captured in pesticide risk assessments, but recent research has demonstrated a need to better understand the sub-lethal impacts of pesticides along with the physiological underpinnings of such impacts.

Researchers have observed that pesticides can disrupt physiological processes that are unrelated to the intended modes of action. Pesticide exposure, for example, can result in oxidative stress [16,20] and apoptosis [18,21] in bees, which is different than the intended modes of action such as disrupting neurological functions (organophosphates and neonicotinoids) or axonic membrane functions (synthetic pyrethroids). If the production of reactive oxygen (ROS) and nitrogen (RNS) species exceeds the body’s natural anti-oxidative defense mechanisms [22,23], then oxidative stress occurs in a number of organisms, including honey bees [24]. An imbalance of excess ROS/RNS-producing environmental factors [25] results in oxidative damage to the organisms. Severe physiological impairments and damage to important macromolecular structures may arise, thereby disrupting vital biological processes [26,27]. Increased systemic oxidative stress, generated by higher levels of ROS/RNS, can play an important role in pesticide toxicity [28–30]. High ROS/RNS levels are also indicative of an onset of apoptosis [30–32]. Apoptosis is a form of programmed cell death [33]. The sublethal impacts of pesticide-induced apoptosis in bees is a serious concern, with effects ranging from midgut cellular lesions to neuronal degeneration in brains [18,33,34]. Caspase-3, a crucial protein in the apoptotic pathway in all organisms, is considered an important marker of apoptotic onset [35,36]. Increased levels of both ROS and caspase-3 are thus indicative of physiological stress. The negative impacts of oxidative stress and apoptosis on individual bees may have cascading effects on the entire colony. For example: oxidative stress may induce precocious foraging (rapid behavioral maturation) leading to colony decline [37].

Our study focuses on the physiological effects of field application rates of two new insecticides that have been developed with an intent to reduce negative effects on bees, when applied to bee-attractive crops. Flupyradifurone, the active ingredient in Sivanto™ (Bayer Crop Science, Germany), and sulfoxaflor, the active ingredient in Transform® (Corteva Agriscience, United States), are two insecticides recently registered for use in the United States [38,39] and are developed to be more compatible with bee health. While Sivanto™ has no restrictions for application during bloom, Transform® has a relatively short residual time on petals and leaf-material and can be legally used during bloom “when managed bees and native pollinators are least active, e.g., 2 hours prior to sunset or when the temperature is below 50˚F at the site of application” [40]. Both active ingredients are systemic, and act as agonists of the nicotinic acetylcholine receptor (nAChR). While both products are considered acutely toxic at field-applied rates when consumed by bees, Sivanto™ is different in that the US labels indicate the product lacks toxicity through contact exposure.

The present study investigated if field application rates of Sivanto™ and Transform® result in sub-lethal physiological effects on honey bees, which extend beyond the recognized mode-of-action of these products. Understanding the lethal and sublethal impacts of Sivanto™, upon contact exposures, is particularly important since its label indicates that the product lacks toxicity through contact exposures. We investigated these questions in laboratory cage experiments, where newly emerged adult honey bees were directly sprayed with field application
rates recommended on the product labels of Transform™ and Sivanto™. We examined survival at 6 hours (lethal effects) and 10 days (sub-lethal effects) in two separate contact exposure experiments. Since pesticide exposures also have the potential to alter food consumption in treated bees [19,41] and affect carbohydrate metabolism, we also measured sugar syrup consumptions. Additionally, physiological impacts were assessed by estimating both oxidative stress (ROS/RNS assay) and apoptosis (caspase-3 protein activity) in the treatment groups.

Methods

Sampling of honey bees for the study

All experiments were conducted in May 2019. The honey bees for the experiments were obtained from six colonies (headed by open mated sister-queens to ensure genetic similarities) at the Oregon State University apiaries in Corvallis, Oregon, USA. The sister-queens were purchased from Jackie Park-Burris Inc. (CA, USA). All six colonies were strong and healthy with low Varroa levels and no visible symptoms of any brood disease. Three frames of capped brood from each colony (18 frames in total) were placed in cardboard nucleus hive boxes and placed in an incubator (Percival Intellus I-36VL, Percival Scientific Inc., USA) overnight at 33.5˚C and 55% RH, allowing bees to emerge. The following day, newly emerged bees from all frames were pooled and gently hand-mixed to ensure a homogenized population. Then, 150 bees were randomly allocated to each replicate cage of each treatment group (see below).

Cylindrical cages, constructed with 0.3175cm hardware cloth and approximately 25.4cm tall and 12.7cm in diameter, were used for this study [42]. Water and 40% sugar syrup (w/v) were provided to bees ad libitum via inverted glass vials on the top of all cages. Evaporation control cages (cages with only sugar syrup and water, and no honey bees) were also included in the study to account for the effects of evaporation. We conducted two separate contact exposure experiments (see below); honey bees for both these experiments were collected and housed as described above.

Experimental design and pesticide exposures

Each of the three experimental groups—control (no pesticide exposure), Sivanto™ (flupyradifurone) and Transform™ (sulfoxaflor)—contained four replicate cages. Sivanto™ Prime (Bayer, active ingredient: Flupyradifurone, 17.09%) and Transform™ WG (Corteva Agriscience, USA, active ingredient: Sulfoxaflor, 50%) were purchased commercially. Each insecticide was mixed at field application rates: (a) Flupyradifurone: 218 μl Sivanto™ Prime in 20 ml water (Sivanto™ Prime package label: 14 fl oz Ac⁻¹ having spray volume 10 gallons Ac⁻¹) and (b) Sulfoxaflor: 33.702 mg Transform™ WG in 20 ml water (Transform™ WG package label: 2.25 oz Ac⁻¹ having spray volume 10 gallons Ac⁻¹). The concentrations of Sivanto™ used in this study represent recommended high rates (in the rate range 7 to 14 fl oz Ac⁻¹) in crops such as citrus fruits and apples to manage pests such as aphids and psyllids (Sivanto™ Prime label). Similarly, the concentrations of Transform™ used in our study represent recommended high rates (in the rate range 0.75 to 2.75 oz Ac⁻¹) for controlling aphids, psyllids and whiteflies in crops such as citrus, apples and other fruiting vegetable crops (Transform™ WG label).

To expose experimental bees to the pesticides, all 150 bees in a given cage were immobilized with carbon dioxide (30 second exposure at 15psi; AirGas, Albany OR) and spread out in a single, even layer on a 140-mm plastic petri dish. Honey bees received a single dose of pesticide via a Potter Spray Tower (Burkard Manufacturing Co. Ltd., England) [43,44]. Each Petri dish of anesthetized bees was placed on the Potter spray tower stage and sprayed with 2 ml of the appropriate insecticide solution or water (for cages in the control group), using an air pressure of 47 kPa (6.8 psi) and a spray distance of 22 cm [44,45]. The bees were then transferred to a
clean Petri dish and placed back in their designated cage. The Potter sprayer delivered the pesticide solution as a fine mist, which formed a uniform layer on the bees, without creating any visible droplets.

Two contact exposure experiments were conducted: one for 6 hours, and another for 10 days. The short-term 6-hours experiment was initially conducted to understand the lethal toxicity of both pesticides to honey bees. Survival and physiological responses were compared between the treatment groups. Since a majority of the honey bees in the Transform™-exposed group died within six hours in the shorter contact exposure experiment, the Transform™ group was excluded from the second, longer 10-day contact exposure experiment. The second experiment was conducted to assess the lethal (mortality observed at the endpoint of the experiment) and the sublethal impacts (physiological alterations observed in the surviving honey bees) of acute contact exposure of Sivanto™ on honey bee physiology. A control group (bees sprayed with only water), was included in both (6-hours and 10-day) experiments.

Survival analyses
Honey bee mortality in each cage was recorded every hour for the 6-hour experiment and daily for the 10-day experiment. Total mortality was calculated at the end of the experiment for each replicate cage in all treatment and control groups. Kaplan-Meier Log Rank Survival analysis was performed [16,46,47] using GraphPad Prism Version 8.0.2 (GraphPad Software, San Diego, USA). Survival curves were calculated for a period up to 6 hours for the shorter contact exposure experiment and 10 days for the longer contact exposure experiment.

Consumption analyses
In the 6-hour experiment, water and sugar syrup consumption were recorded at the end of the experiment. Each glass vial was initially filled with 30 ml of sugar syrup or water and the volume of liquid remaining in each vial was recorded at the end of the experiment, and subtracted from the initial volume given (30 ml). The volume of liquid lost through evaporation (from vials in evaporation control cages) was subtracted from the volume lost in vials from honey bee-containing cages. This was then adjusted to the number of live honey bees at the end of the experiment to calculate consumption per bee. Similar estimations for water and sugar syrup consumptions were made each day for the 10-day experiment, where water and sugar syrups were replaced daily.

At the end of both the 6-hours and 10-days experiments, five live honey bees were sampled from each replicate cage in the Sivanto™ and control groups for estimating oxidative stress. Five live honey bees were also collected from those cages for estimating caspase-3 activity. Likewise, we sampled five live honey bees for each assay (10 live bees in total) from each of the Transform™-treatment group replicate cages at the end of the 6-hour experiment. All the sampled bees were immediately frozen at -80 °C until time of analysis.

Activity assays of reactive oxygen species (an indicator of pesticide induced oxidative stress)
As mentioned before, five honey bees from each replicate cage, for every treatment group, were pooled together. Whole honey bees were homogenized in phosphate buffered saline (PBS, Sigma Aldrich, USA; 10 mM phosphate, 2.7 mM potassium chloride and 137 mM sodium chloride, pH 7.4) using a Precellys 24 tissue homogenizer (Bertin Instruments, Rockville, USA) at 6000 rpm for three cycles of 30 seconds per cycle. The tissue homogenates were then centrifuged at 10,000g for 5 minutes at 4°C (Eppendorf model 5430R, Eppendorf, USA)
and supernatants were immediately assayed for reactive oxygen/nitrogen species as described below.

Oxidative stress was measured by assaying total reactive oxygen species (ROS) and total reactive nitrogen species (RNS) activities using OxiSelect™ In vitro ROS/RNS Assay Kit—Green Fluorescence (Cell Biolabs. Inc., San Diego, USA) [48,49] using the manufacturer’s microplate assay protocols. Briefly, the assay principle is based on oxidizing a fluorogenic probe (2',7'-dichlorodihydrofluorescin DiOxyQ) to a fluorescent product (2',7'-dichlorodihydrofluorescein) in the presence of ROS and RNS. The fluorescence intensity is proportional to the amount of ROS/RNS present in the sample. Fluorescence was measured at 480 nm excitation and 530 nm emission wavelengths by a microplate reader (Biotek Synergy 2, BioTek Instruments, USA) using a hydrogen peroxide standard curve. The pooled tissue homogenates from each replicate cage were assayed in duplicate in a 96-well plate. We used general protein content to normalize the results of oxidative stress assays across all samples, such that total ROS/RNS activity is expressed as μmol μL⁻¹ μg⁻¹ protein. Therefore, protein estimations of the homogenate supernatants were performed using a BCA Assay Kit (Thermo Scientific, USA) following the microplate assay protocol, and the absorbance was measured at 562 nm (BioTek Synergy 2 plate reader, BioTek Instruments, USA) [42].

Caspase-3 protein activity assay (an indicator of cell death/apoptosis)

For estimating caspase-3 activity, five honey bees were sampled and pooled from each of the experimental cages (for all treatments and controls) and homogenized and centrifuged as described above for ROS/RNS assays. The resulting supernatants were immediately assayed. Apoptosis was estimated by detecting the increase in caspase-3 activity using the fluorometric caspase-3 assay (Abcam, Cambridge, USA) following manufacturer’s protocols [50,51]. Sample supernatants were assayed in duplicate and fluorescence was measured at 400 nm excitation and 505 nm emission by a microplate reader (Biotek Synergy 2, BioTek Instruments, USA). As with ROS/RNS assays, caspase-3 activity was normalized using general protein content. Thus, protein content was quantified using a BCA Assay Kit (Thermo Scientific, USA) as described above [42]. Caspase-3 activity is expressed as U μg⁻¹ protein.

Statistical analyses

All statistical analyses were performed using GraphPad Prism Version 8.0.2 (GraphPad Software, San Diego, USA). Shapiro-Wilk test for normality was conducted for every dataset. For comparing two groups, two tailed t-tests were performed for data that were normally distributed and Mann Whitney U test for non-parametric tests. For parametric tests, multiple comparisons were performed using One-Way ANOVA with Tukey’s Posy Hoc comparisons. Kruskal-Wallis test with Dunn’s multiple comparisons were performed for data that were not normally distributed. The data for caspase-3 was log transformed. Means are presented as means ± standard errors of means.

Results

Survival

Six-hour contact exposure experiment. The survival curves were significantly different among the experimental groups in the 6-hours contact exposure experiment (Kaplan Meier, Log Rank test $\chi^2 = 4719$, df = 2 and p < 0.001) with the highest bee survival recorded in the control group, followed by the Sivanto™ and Transform™ groups (Fig 1a).
Ten-day contact exposure experiment. Log-rank test indicated significant differences in bee survival among experimental groups (Kaplan Meier, Log Rank test $\chi^2 = 192.6$, df = 1 and $p < 0.001$). While more bees survived in the control group compared to the Sivanto™ group, this difference became pronounced by the end of the experiment when control cages had 70% survival compared to 45% in the Sivanto™ group (Fig 1b).

Consumption of water and sugar syrup

Six-hour contact exposure experiment. Consumption of water in the 6-hour experiment was significantly different between the treatment groups (One-way ANOVA with Tukey’s Post Hoc comparisons, $F_{(2,9)} = 19.28$, $p < 0.001$) (Fig 2a). Control, Sivanto™ and Transform® treatment groups consumed 23.14±3.23, 26.25±4.39 and 75.78±10.29 μL bee$^{-1}$ cage$^{-1}$ respectively. However, sugar syrup consumption was not significantly different (One-way ANOVA with Tukey’s Post Hoc comparisons, $F_{(2,9)} = 1.706$, $p = 0.2354$) with control bees consuming 24.14±0.84 μL bee$^{-1}$ cage$^{-1}$, Sivanto™-exposed bees consuming 33.32±4.39 μL bee$^{-1}$ cage$^{-1}$ and Transform®-exposed bees consuming 35.79±6.80 μL bee$^{-1}$ cage$^{-1}$ of sugar syrup (Fig 2a).

Ten-day contact exposure experiment. For the 10-day experiment, water consumption was not significantly different between the experimental groups ($t = 1.962$, df = 6, $p = 0.0974$), even though higher consumption was observed in the Sivanto™ group (159.32±34.43 μL bee$^{-1}$ day$^{-1}$ cage$^{-1}$), when compared with the control group (89.98±7.90 μL bee$^{-1}$ day$^{-1}$ cage$^{-1}$) (Fig 2b). Sugar syrup consumption differed significantly between experimental groups ($t = 4.766$, df = 6, $p < 0.05$), with control and Sivanto™-groups consuming sugar syrup as follows: 73.36±2.16 μL bee$^{-1}$ day$^{-1}$ cage$^{-1}$ and 81.84±3.86 μL bee$^{-1}$ day$^{-1}$ cage$^{-1}$ respectively (Fig 2b).

Pesticide induced oxidative stress in honey bees

Six-hour contact exposure experiment. The total ROS/RNS activity was significantly different between the treatment groups (One-way ANOVA with Tukey’s Post Hoc comparisons, $F_{(2,21)} = 23.56$, $p < 0.001$) (Fig 3a). Both the Sivanto™ exposed honey bees and the Transform® exposed honey bees expressed significantly higher ROS/RNS activities, when compared with the control group (2.19±0.24 and 2.47±0.05 fold increases respectively). There was no significant difference in ROS/RNS activity between Sivanto™ and Transform® treatment groups ($p = 0.5615$). The mean ROS/RNS activity values for control, Sivanto™ and Transform®...
treatment groups were 0.48±0.01 μmol μL⁻¹ μg⁻¹ protein, 1.07±0.13 μmol μL⁻¹ μg⁻¹ protein and 1.18±0.023 μmol μL⁻¹ μg⁻¹ protein respectively (Fig 3a).

**Ten-day contact exposure experiment.** Total ROS/RNS activity significantly differed between the treatment groups (Mann Whitney U = 4, p < 0.05) (Fig 3b). The total ROS/RNS activity was significantly higher (1.975±0.33 fold increase) in the Sivanto™ treatment group when compared to the control. The mean ROS/RNS activities were 0.35±0.05 μmol μL⁻¹ μg⁻¹ protein and 0.58±0.03 μmol μL⁻¹ μg⁻¹ protein in control and Sivanto™ groups respectively (Fig 3b).
Fig 3. Mean ROS/RNS concentration in bees that were exposed (contact exposure) to Transform® (active ingredient sulfoxaflor), Sivanto™ (active ingredient flupyradifurone) and water (Control) in the (a) 6-hour and (b) 10-day contact exposure experiments. Error bars indicate SEM. Different alphabets indicate statistical significance at $p < 0.05$, as analyzed by Tukey’s Post Hoc tests. * indicates statistical significance at $p < 0.05$.

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Pesticide induced cell death (apoptosis) in honey bees

**Six-hour contact exposure experiment.** The caspase-3 activity was significantly different between treatment groups (Kruskal-Wallis test $H = 12.26, p < 0.05$) (Fig 4a). The bees in the Transform® treatment group had significantly higher caspase-3 activity ($1.31 \pm 0.09$ fold higher) when compared to bees in the control group (Dunn’s multiple comparisons, $z = 3.5, p < 0.05$) (Fig 4a). No significant differences were observed in caspase-3 activities between Sivanto™ exposed bees and bees in the control group (Dunn’s multiple comparisons, $z = 1.803, p = 0.2141$). Similarly, the caspase-3 activity was not significantly different between Sivanto™ and Transform® exposed bees (Dunn’s multiple comparisons, $z = 1.697, p = 0.2691$). The mean caspase-3 activity for the control, Sivanto™ and Transform® treatment groups were $3167.55 \pm 0.03 \text{ U} \mu\text{g}^{-1} \text{ protein}$, $3633.26 \pm 0.01 \text{ U} \mu\text{g}^{-1} \text{ protein}$ and $4105.023 \pm 0.02 \text{ U} \mu\text{g}^{-1} \text{ protein}$ respectively (Fig 4a).

**Ten-day contact exposure experiment.** Caspase-3 activity was significantly different between treatment groups ($t = 4.241, \text{df} = 14, p < 0.001$) (Fig 4b). Caspase-3 activity was $1.35 \pm 0.09$ times higher in the Sivanto™-exposed group than in the control group (Fig 4b). Mean caspase-3 activities were $3302.83 \pm 0.02 \text{ U} \mu\text{g}^{-1} \text{ protein}$ and $4406.45 \pm 0.02 \text{ U} \mu\text{g}^{-1} \text{ protein}$ in control and Sivanto™ groups respectively (Fig 4b).

**Discussion**

This study reports, for the first time, increased oxidative stress and onset of apoptosis in honey bees exposed to two relatively recently registered insecticides—Transform® and Sivanto™. Our findings provide new insights on sub-lethal and lethal impacts of contact exposures of field application rates of these two pesticides (formulated products) on honey bees. Both bee mortality and parameters related to physiological stress in honey bees were measured for a comprehensive understanding of potential negative impacts of these pesticides.

A majority of the honey bees exposed to Transform® died within the six hours after initiation of the experiment, which confirms severe toxicity of Transform® to bees when exposed directly to field application rates recommended on the label. Further, our findings also confirm that Sivanto™ is not directly lethal to honey bees following contact exposure. Our 10-day survival results, however, revealed evidence that field-rates of Sivanto™ may reduce adult survival. This suggests that even though Sivanto™ is apparently less toxic than Transform®, it may reduce honey bee longevity. This is further evident in the survival analyses, where honey bees in the control groups exhibited the highest survival in both contact exposure studies. Consequently, we provide data that contact exposure routes, in addition to known oral routes of exposure, may be an additional source of acute toxicity of Sivanto™. Notably, our work investigated the direct application of Sivanto™ to bees, not the contact of bees with field-weathered residues of the formulated product on leaves and flower petals. Future studies should investigate whether similar effects can be replicated when bees contact residues found on plants following treatments.

Consumption of sugar syrup was not significantly different in the 6-hour experiment. However, Sivanto™ exposed honey bees consumed significantly more sugar syrup than the control honey bees over the 10-day period. Pesticide exposures are known to alter sugar syrup consumption in treated bees [19,41]. The energetics of carbohydrate metabolism and regulation may be responsible for such consumption differences. Pesticides are reported to impact glucose and glycogen levels in bee tissues, altering various carbohydrate signaling and molecular pathways [19]. Examining these mechanisms was beyond the scope of this study. The differential syrup consumption we observed, however, further indicates that pesticides may influence ...
Fig 4. Mean caspase-3 activity in bees exposed to Transform™ (active ingredient sulfoxaflor), Sivanto™ (active ingredient flupyradifurone) and water (Control) in the (a) 6-hours and (b) 10-days contact exposure studies. Error bars indicate SEM. Different alphabets indicate statistical significance at p < 0.05, as analyzed by Tukey’s Post Hoc tests. * indicates statistical significance at p < 0.05.

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bee metabolic function. For similar metabolic reasons, water consumption may also have differed between the treatment groups in both contact exposure studies.

The present study further illustrates pesticide induced oxidative stress in honey bees. In the shorter 6-hour experiment, Transform™ exposed honey bees exhibited the highest oxidative stress (significantly higher than control) when compared with honey bees in other treatment groups. In both Sivanto™ contact groups, ROS/RNS was significantly higher than the control group. However, it should be noted that mean ROS/RNS concentrations were less at the end of the longer 10-day experiment, when compared with the shorter 6-hour exposure experiment. This may indicate an immediate oxidative stress in the Sivanto™ groups during the shorter study, which may have been physiologically mitigated over the 10-day duration as seen in the longer contact exposure study. Even though Apis mellifera has fewer detoxification genes compared to other insects [52], they are known to elevate antioxidant enzymes (for example: catalase, superoxide dismutase, xanthine oxidase etc.) [16] or develop tolerance to mitigate such physiological stressors in their bodies to offset pesticide induced oxidative stress [53]. Both flupyradifurone and sulfoxaflor target insect nAChR and other molecular targets may be potentially affected as well [54]. A number of neonicotinoids are known to increase production of reactive oxygen species in honey bees, but direct evidence of increased oxidative stress in bees, upon contact exposures to Sivanto™ (active ingredient flupyradifurone) and Transform™ (active ingredient sulfoxaflor) at field-application rates, was lacking [54].

Initiation of cell death and apoptosis was also evident in both contact exposure experiments. Caspase-3 is an important marker of apoptosis, and has been reported to increase under pesticide stress in all organisms, including honey bees [30,33,35,36]. Significantly increased activity of caspase-3 in Transform™-treated honey bees in the 6-hour contact exposure experiment (compared to control), indicated physiological damage that may occur from this pesticide at field application rates. In bees exposed to Sivanto™, caspase-3 activity was significantly elevated (compared to control) at the end of the 10-day period (longer study), and not after 6 hours (short duration study), indicating the sublethal effects of this pesticide on bees in the long term. We speculate that apoptosis was potentially initiated in the bees that survived until the end of the experiment, but was not discerned during the short term (after six hours). This finding is important when assessing impacts of such pesticides in the field over a longer duration of time.

Physiological impacts of pesticides, for example oxidative stress and apoptosis, can render individual honey bees incapable of performing their tasks smoothly, thereby affecting the colony performance as well. Previous studies have shown that different species of honey bees exhibit different anti-oxidant enzyme responses to field-realistic oral exposures pesticides (Apis cerana and Apis dorsata) [16], as well as varying conditions of temperature and humidity (Apis cerana and Apis mellifera) [55]. Further, pesticide-induced oxidative stress, coupled with other stressors, can reduce longevity in honey bees [56]. Thus, reduced longevity resulting from oxidative stress could negatively affect colony population and ultimately compromise colony fitness. Previous studies have also asserted the need to conduct similar long-term monitoring for pesticide exposures [57,58], which may help assess the long-term impacts of such exposures on colony health. Even though the present study uses total ROS/RNS and caspase-3 activity assays as indicators for physiological stress, future studies should look into other proteins of the apoptotic pathways and antioxidant enzymes.

Recent studies on flupyradifurone [8,54,59] and sulfoxaflor [60,61] have demonstrated lethal and sublethal impacts of these pesticides in terms of survival, cognition and abnormal behavior in honey bees. However, to our knowledge, the physiological underpinnings of these impacts due to direct contact exposures are not known. Our study has attempted to address this gap in knowledge. Further, our findings also suggest the need to revisit the potential
impacts of contact exposure of Sivanto™; given the sub-lethal impacts and reduced longevity in honey bees observed in our long-term contact exposure experiment (10-day experiment). Future research should focus on the impacts of such field-relevant contact exposures across multiple population cohorts at different seasonal time points. With the recent Environmental Protection Agency approval for use of both flupyradifurone and sulfoxaflor, and with the growing concern regarding pollinator health, it is important to better understand any potential negative impacts (especially sub-lethal) of these pesticides on bees.

Supporting information
S1 Data.
(XLSX)

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References
1. Frazier M, Mullin C, Frazier J, Ashcraft S. What have pesticides got to do with it? Am Bee J. 2008; 148: 521–523.
2. Winfree R., Aguilar R., Vázquez DP, LeBuhn G, Aizen MA. A meta-analysis of bees’ responses to anthropogenic disturbance. Ecology. 2009; 90: 2068–2076. https://doi.org/10.1890/08-1245.1 PMID: 19739369
3. Henry N, Béguin M, Requier F, Rollin O, Odoux J.-F, Aupinel P, et al. A common pesticide decreases foraging success and survival in honey bees. Science 2012; 336: 348–350. https://doi.org/10.1126/science.1215039 PMID: 22461498
4. Pettis JS, vanEngelsdorp D, Johnson J, Dively G. Pesticide exposure in honey bees results in increased levels of the gut pathogen *Nosema*. Naturwissenschaften. 2012; 99: 153–158. https://doi.org/10.1007/s00114-011-0881-1 PMID: 22246149

5. Palmer MJ, Moffat C, Saranzedwa N, Harvey J, Wright GA, Connolly CN. Cholinergic pesticides cause mushroom body neuronal inactivation in honeybees. Nat Commun. 2013; 4: 634.

6. Chakrabarti P, Rana S, Bandopadhyay S, Naik DG, Sarkar S, Basu P. Field populations of native Indian honey bees from pesticide intensive agricultural landscape show signs of impaired olfaction. Sci Rep. 2015; 5: 12504. https://doi.org/10.1038/srep12504 PMID: 26212690

7. Chakrabarti P, Sarkar S, Basu P. Pesticide induced visual abnormalities in Asian honey bees (*Apis cerara*) in intensive agricultural landscapes. Chemosphere. 2019; 230: 51–58. https://doi.org/10.1016/j.chemosphere.2019.05.050 PMID: 31102871

8. Tosi S, Nieh JC. Lethal and sublethal synergistic effects of a new systemic pesticide, flupyradifurone (Sivanto), on honeybees. Proc R Soc B. 2019; 286: 20190433. https://doi.org/10.1098/rspb.2019.0433 PMID: 30966981

9. Whitehorn PR, O’Connor S, Wackers FL, Goulson D. Neonicotinoid pesticide reduces bumble bee colony growth and queen production. Science. 2012; 336: 351–352. https://doi.org/10.1126/science.1215025 PMID: 22461500

10. Belzunces LP, Tchemitchian S, Brunet JL. Neural effects of insecticides in the honey bee. Apidologie 2012; 43: 348–370.

11. Williams GR, Troxler A, Retschnig G, Roth K, Yanez O, Shutter D, et al. Neonicotinoid pesticides severely affect honey bee queens. Sci Rep. 2015; 5: 14621. https://doi.org/10.1038/srep14621 PMID: 26459072

12. Wu-Smart J, Spivak M. Sub-lethal effects of dietary neonicotinoid insecticide exposure on honey bee queen fecundity and colony development. Sci Rep. 2016; 6: 32108. https://doi.org/10.1038/srep32108 PMID: 27562025

13. Straub L, Villamar-Bouza L, Bruckner S, Chantawannakul P, Gauthier L, Khongphinthunjong K., et al. Neonicotinoid insecticides can serve as inadvertent insect contraceptives. Proc R Soc B Biol Sci. 2016; 283: 1835.

14. Tosi S, Burgio G, Nieh JC. A common neonicotinoid pesticide, thiamethoxam, impairs honey bee flight ability. Sci Rep. 2019; 7: 1201.

15. Muth F, Leonard AS. A neonicotinoid pesticide impairs foraging, but not learning, in free-flying bumblebees. Sci Rep. 2019; 9: 4764. https://doi.org/10.1038/s41598-019-39701-5 PMID: 30886154

16. Chakrabarti P, Rana S, Smith B, Sarkar S, Basu P. Pesticide induced oxidative stress in laboratory and field populations of native honey bees along intensive agricultural landscapes in two Eastern Indian states. Apidologie 2015; 46: 107e129.

17. Simone-Finstrom M, Li-Byarlay H, Huang MH, Strand MK, Rueppell O, Tarpy DR. Migratory management and environmental conditions affect lifespan and oxidative stress in honey bees. Sci Rep. 2016; 6: 32023. https://doi.org/10.1038/srep32023 PMID: 27554200

18. Gregorc A, Alburaki M, Rinderer N, Sampson B, Knight PR, Karim S, et al. Effects of coumaphos and imidacloprid on honey bee (Hymenoptera: Apidae) lifespan and antioxidant gene regulations in laboratory experiments. Sci Rep. 2018; 8: 15003 (2018). https://doi.org/10.1038/s41598-018-33348-4 PMID: 30301926

19. Cook SC. Compound and Dose-Dependent Effects of Two Neonicotinoid Pesticides on Honey Bee (*Apis mellifera*) Metabolic Physiology. Insects 2019; 10: 18. https://doi.org/10.3390/insects1010018 PMID: 30626039

20. Alburaki M, Steckel SJ, Chen D, McDermott E, Weiss M, Skinner JA, et al. Landscape and pesticide effects on honey bees: forager survival and expression of acetylecholinesterase and brain oxidative genes. Apidologie. 2017; 48: 556–571.

21. Johnson RM, Ellis MD, Mullin C, Frazier M. Pesticides and honey bee toxicity—USA. Apidologie 2010; 41: 312–331.

22. Velki M, Kodrik D, Večera J, Hackenberger BK, Socha R Oxidative stress elicited by insecticides: a role for the adipokinet hormone. Gen Comp Endocrinol. 2011; 172: 77–84. https://doi.org/10.1016/j.ygece.2010.12.009 PMID: 21185291

23. Nishida Y. The chemical process of oxidative stress by copper (II) and iron (III) ions in several neurodegenerative disorders. Monatsh Chem. 2011; 142: 375–384.

24. Paynter E, Millar A, Welch M, Bae-Imhoof B, Cao D, Baer B. Insights into the molecular basis of long-term storage and survival of sperm in the honeybee (*Apis mellifera*). Sci Rep. 2017; 7: 40236. https://doi.org/10.1038/srep40236 PMID: 28091518
32. Ali D, Tripathi A, Ali HA, Shahi Y, Mishra KK, Alarifi S, et al. ROS-dependent Bax/Bcl2 and caspase 3
31. Redza-Dutordoir M, Averill-Bates DA. Activation of apoptosis signalling pathways by reactive oxygen
33. Wu YY, Zhou T, Wang Q, Dai PL, Xu SF, Jia HR, et al. Programmed Cell Death in the Honey Bee (Apis mellifera) (Hymenoptera: Apidae) Worker Brain Induced by Imidacloprid. J Econ Entomol. 2015; 108: 1486–1494. https://doi.org/10.1093/jee/tov146 PMID: 26470287
34. Jablonska-Trypuć A. Pesticides as Inducers of Oxidative Stress. React Oxyg Species. 2017; 3: 96–110.
35. Porter AG, Jänické RU. Emerging roles of caspase-3 in apoptosis. Cell Death Differ. 1999; 6: 99–104. https://doi.org/10.1038/sj.cdd.4400476 PMID: 10200555
36. McIlnaw DR, Berger T, Mak TW. Caspase functions in cell death and diseases. Cold Spring Harb Perspect Biol. 2013; 5: a008656. https://doi.org/10.1101/cshperspect.a008656 PMID: 23545416
37. Perry CJ, Savik E, Myerscough MR, Barron AB. Rapid behavioral maturation accelerates failure of stressed honey bee colonies. PNAS 2015; 112: 3427–3432. https://doi.org/10.1073/pnas.1422089112 PMID: 25675508
38. U.S. Environmental Protection Agency, Pesticide Product Label, notice of pesticide registration of Sulfoxaflor Technical (EPA Reg. No. 62719–631) 2013. http://www3.epa.gov/pesticides/chem_search/pplis/062719-00631-20130506.pdf.
39. U.S. Environmental Protection Agency, Pesticide Product Label, notice of pesticide registration of Fluopyradifurone TC (EPA Reg. No. 264–1143) 2015. http://www3.epa.gov/pesticides/chem_search/pplis/000264-01143-20150115.pdf.
40. U.S. Environmental Protection Agency, 2019. https://www3.epa.gov/pesticides/chem_search/pplis/062719-00625-20191127.pdf.
41. Alburaki M, Karim S, Lamour K, Adamczyk J, Stewart SD. RNA-seq reveals disruption of gene regulation when honey bees are caged and deprived of hive conditions. J Exp Biol. 2019; 222: jeb207761. https://doi.org/10.1242/jeb.207761 PMID: 31413101
42. Chakrabarti P, Lucas HM, Sagili RR. Evaluating Effects of a Critical Micronutrient (24-Methylenecholesterol) on Honey Bee Physiology. Ann Entomol Soc Am. 2019; 1–7. https://doi.org/10.1093/aes/aesz067 PMID: 32410742
43. Potter C, Way MJ. Precision spraying. In: Shepard HH, editor. Methods of Testing Chemicals on Insects. Minneapolis: Burgess Publishing Company; 1958. pp. 154.
44. Zhu YC, Adamczyk J, Rinderer T, Yao J, Danko R, Luttrell R, et al. Spray Toxicity and Risk Potential of 42 Commonly Used Formulations of Row Crop Pesticides to Adult Honey Bees (Hymenoptera: Apidae). J Econ Entomol. 2015; 108: 2640–2647. https://doi.org/10.1093/jee/tov269 PMID: 26352753
45. Zhu YC, Yao J, Adamczyk J, Luttrell R. Feeding toxicity and impact of imidacloprid formulation and mixtures with six representative pesticides at residue concentrations on honey bee physiology (Apis mellifera). PLoS ONE. 2017; 12: e0178421. https://doi.org/10.1371/journal.pone.0178421 PMID: 28591204
46. Klein JP, Moeschberger ML. Survival Analysis. Techniques for Censored and Truncated Data. 2nd ed. New York: Springer Publishers; 2003.
47. Saingamsook J, Yanola J, Lumjuan N, Walton C, Somboon P. Investigation of Relative Development and Reproductivity Fitness Cost in Three Insecticide-Resistant Strains of Aedes aegypti from Thailand. Insects 2019; 10: 265. https://doi.org/10.3390/insects10090265 PMID: 31443487

48. Perraulta JR, Baumanb KD, Greenan TM, Blum PC. Maternal transfer and sublethal immune system effects of brevetoxin exposure in nesting loggerhead sea turtles (Caretta caretta) from western Florida. Aquat Toxicol. 2016; 180: 131–140. https://doi.org/10.1016/j.aquatox.2016.09.020 PMID: 27716578

49. Zhelev Z, Ivanova D, Bakalova R, Aoki I, Higashi T. Synergistic Cytotoxicity of Melatonin and New-generation Anticancer Drugs Against Leukemia Lymphocytes But Not Normal Lymphocytes. Anticancer Res. 2017; 37: 149–160. https://doi.org/10.21873/anticancerres.11300 PMID: 28011485

50. Lin SF, Lin JD, Hsueh C, Chou TC, Wong RJ. Activity of roniciclib in medullary thyroid cancer. Oncotarget. 2018; 9: 28030–28041. https://doi.org/10.18632/oncotarget.25555 PMID: 29963260

51. Droho S, Keener ME, Mueller NH. Changes in function but not oligomeric size are associated with aB-crystallin lysine substitution. Biochem Biophys Rep. 2018; 14: 1–6. https://doi.org/10.1016/j.bbrep.2018.03.001 PMID: 29872727

52. Claudianos C, Ranson H, Johnson RM, Biswas S, Schuler MA, Berenbaum MR, et al. A deficit of detoxification enzymes: pesticide sensitivity and environmental response in the honeybee. Insect Mol Biol. 2006; 15: 615–636. https://doi.org/10.1111/j.1365-2583.2006.00672.x PMID: 17069637

53. Li-Byarlay H, Huang MH, Simone-Finstrom M, Strand MK, Tarpy DR, Ruepell O. Honey Bee (Apis mellifera) Drones Survive Oxidative Stress due to Increased Tolerance instead of Avoidance or Repair of Oxidative Damage. Exp Gerontol. 2016; 83: 15–21. https://doi.org/10.1016/j.exger.2016.07.003 PMID: 27422326

54. Hesselbach H, Scheiner R. The novel pesticide flupyradifurone (Sivanto) affects honeybee motor abilities. Ecotoxicology. 2019; 28: 354–366. https://doi.org/10.1007/s10646-019-02028-y PMID: 30826953

55. Li X, Ma W, Shen J, Long D, Feng Y, Su W, et al. Tolerance and response of two honeybee species Apis cerana and Apis mellifera to high temperature and relative humidity. PLoS ONE 14: e0217921. https://doi.org/10.1371/journal.pone.0217921 PMID: 31170259

56. Guilin L, Hang Z, Zhenguo L, Hongfang W, Baohua X, Xinggi G. The Wisdom of Honeybee Defenses Against Environmental Stresses. Front Microbiol. 2018; 9: 722. https://doi.org/10.3389/fmicb.2018.00722 PMID: 29765357

57. Williamson SM, Baker DD, Wright GA. Acute exposure to a sublethal dose of imidacloprid and coumaphos enhances olfactory learning and memory in the honeybee Apis mellifera. Invertebr Neurosci. 2013; 13: 63–70.

58. Dolezal AG, Carrillo-Tripp J, Miller WA, Bonning BC, Toth AL. Pollen Contaminated With Field-Relevant Levels of Cyhalothrin Affects Honey Bee Survival, Nutritional Physiology, and Pollen Consumption Behavior. Journal of Econ Entomol. 2016; 109: 41–48.

59. Hesselbach H, Scheiner R. Effects of the novel pesticide flupyradifurone (Sivanto) on honeybee taste and cognition 2018. Sci Rep. 2018; 8: 4954. https://doi.org/10.1038/s41598-018-23200-0 PMID: 29563522

60. Zhu YC, Yao J, Adamczyk J, Luttrell R. Synergistic toxicity and physiological impact of imidacloprid alone and binary mixtures with seven representative pesticides on honey bee (Apis mellifera). PLoS ONE. 2017; 12: e0176837. https://doi.org/10.1371/journal.pone.0176837 PMID: 28467462

61. Cheng Y, Bu Y, Tan L, Wu W, Li J, Zhou J, et al. A semi-field study to evaluate effects of sulfoxaflor on honey bee (Apis mellifera). Bull Insectology. 2018; 71: 225–233.