In silico approach to investigating the adsorption mechanisms of short chain perfluorinated sulfonic acids and perfluorooctane sulfonic acid on hydrated hematite surface

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\textbf{Abstract}

Short chain perfluorinated sulfonic acids (PFSAs) that were introduced as alternatives for perfluorooctane sulfonic acid (PFOS) have been widely produced and used. However, few studies have investigated the environmental process of short chain PFSAs, and the related adsorption mechanisms still need to be uncovered. The water-oxide interface is one of the major environmental interfaces that plays an important role in affecting the adsorption behaviour and transport potential of the environmental pollutant. In this study, we performed molecular dynamics simulations and quantum chemistry calculations to investigate the adsorption mechanisms of five PFSAs and their adsorption on hydrated hematite surface as well. Different to the vertical configuration reported for PFOS on titanium oxide, all PFSAs share the same adsorption configuration as the long carbon chains parallel to the surface. The formation of hydrogen bonds between F and inter-surface H helps to stabilize the unique configuration. As a result, the sorption capacity increases with increasing C-F chain length. Moreover, both calculated adsorption energy and partial density of states (PDOS) analysis demonstrate a PFSAs adsorption mechanism in between physical and chemical adsorption because the hydrogen bonds formed by the overlap of F\,(p) orbital and H\,(s) orbital are weak intermolecular interactions while the physical adsorption are mainly ascribed to the electrostatic interactions. This massive calculation provides a new insight into the pollutant adsorption behaviour, and in particular, may help to evaluate the environmental influence of pollutants.

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\section{1. Introduction}

Perfluoroalkyl acids are widely used as surfactants in various industrial and consumer products such as lubricants, paints, metal cleaners, cosmetics, and fire-fighting foams (Matyszewska et al., 2007). They have been detected in the environment and drinking water and have been found to be present in humans and a wide range of biota (Giesy and Kannan, 2001; Kannan et al., 2004; Rahman et al., 2014; Thomas, 2006). Due to their persistence, toxicity, bioaccumulative behaviour, and the capacity for long-range transport, they have attracted worldwide attention; this is especially true for perfluorooctane sulfonic acid (PFOS) and perfluorooctane sulfonyl fluoride (POSF)-based chemicals that were listed under Annex B (restriction of production and use) of the UNEP Stockholm Convention in 2009 (UNEP, 2009). Hence, the short chain perfluorinated sulfonic acids (PFSAs, \( C_{n}F_{2n+1}SO_{3}H \)) with fewer than seven perfluorinated carbons and their salts were produced and used as alternative compounds due to their identical functional properties and lower bioaccumulation (Olsen et al., 2009; Wang et al., 2013). Previous studies have reported that...
PFBS was found to be the dominant PFAs in the Rhine River watershed, and drinking water in Germany and the Tangxun Lake (Lange et al., 2007; Moller et al., 2010; Zhou et al., 2013). Therefore, attention should be concentrated on the environmentally related processes of the PFASs in aquatic environment. Despite the increasing production of short chain PFASs and the fact that short chain PFASs have spread worldwide, little is known about their processes in the environment.

Due to the heterogeneity in the environmental interface, aqueous and mineral surface chemical behaviours are the major controlling factors in the fate and transport of organic pollutants. Mineral surfaces that are often used as catalysts and adsorbents play a key role in several environmental processes. Previous studies have shown that mineral surfaces in contact with ground water can adsorb and catalyze reactions of pollutant molecules (Al-Abadleh and Grassian, 2003). Hematite (α-Fe₂O₃) is a common and ubiquitous environmental sorbent and thus plays an important role in contaminant sequestration and chemical speciation in ground-water aquifers and soils (Jr et al., 1999a,b). Studies of environmental processes provide the basis for toxicity and potential bioavailability of contaminants (Jr et al., 1999b), which in turn affect the contaminant sequestration and chemical speciation in groundwater. Li et al. (2012) studied the glycine molecules in the gas phase on rutile TiO₂ (1 1 0) surface in a water solution by molecular dynamics simulations (Li et al., 2012). Notman and Walsh (2009) investigated the interactions of water and the surface of α-quartz by molecular dynamics simulations. Li et al. (2012) studied the glycine molecules in the gas phase on rutile TiO₂ (1 1 0) surface in a water solution by molecular dynamics simulations (Li et al., 2012). The surface was known to hydroxylate at water vapour pressure greater than 10⁻⁴ Torr. It was believed that the surface was fully hydroxylated in aqueous conditions (Liu et al., 1998). We noticed that there several different hydrated hematite (0 0 1) 0.001 surfaces are known (Mason et al., 2009; Souvi et al., 2013; Wasserman et al., 1997), and in the consideration of our model and the ambient environment, the surface model was built symmetrically so that there is no net surface dipole. The construction of hydrated hematite (0 0 0 1) surface is presented detailedly in Supporting Information, Text S1.

2. Computational details and models

2.1. Computational details

All simulations were carried out with the commercial Materials Studio software package (Accelrys Software Inc., 2012). The simulation methodology includes molecular dynamics calculations using the Forcite module and density functional theory calculations using the CASTEP module (Clark et al., 2005).

MD was performed in the periodic canonical ensemble (NVT) using the COMPASS force field, which is one of the first ab initio force field approaches that has been parameterized and validated using the gas-phase and condensed-phase properties of organic and inorganic materials (Sun, 1998; Sun et al., 1998). The minimization was performed using the Smart algorithm that automatically combines appropriate features of the other available methods in a cascade. Simulations were carried out in the NVT ensemble using the Nosé algorithm (Karasawa and lli, 1989; Nosé, 1991, 2006, 1984) for a run time of 4 ns with a 1 fs time step at 298 K. The simulations were initialized with the minimum-energy structures and random initial velocities. The non-bonded interactions were calculated using the Ewald (Hohenberg and Kohn, 2010; Williams, 2007) method.

To obtain a more accurate hematite structure, DFT calculations were carried out with the CASTEP module using the plane wave pseudopotential method (PWP) within the DFT formalism (Kohn and Sham, 1965). For the model of hematite (α-Fe₂O₃), spin polarization should be considered. The detailed parameters of spin direction, spin state and DFT-U are found in previous studies (Martín et al., 2009; Rollmann et al., 2004; Träino et al., 2004). More computational details are presented in Support Information, Text S1.

2.2. Computational models

The geometry optimized α-Fe₂O₃ hexagonal unit cell \(a = b = 5.02892 \text{ Å}, c = 13.7956 \text{ Å}, \alpha = \beta = 90^\circ, \gamma = 120^\circ\) was in good agreement with the experimental data \(a = b = 5.038 \text{ Å}, c = 13.772 \text{ Å} \) (Blake et al., 1966); \(a = b = 5.035 \text{ Å}, c = 13.747 \text{ Å} \) (Finger and Hazen, 1980). The relaxed crystal was used to cleave the hematite (0 0 0 1) surface that was frequently observed on the naturally grown crystals and that is thermodynamically stable (Wang et al., 1998). Previous studies indicated that α-Fe₂O₃ (0 0 0 1) is Fe-terminated and that the highly relaxed single-layer Fe-terminated Fe-O₃-Re surface is the most stable configuration (Alvarez-Ramírez et al., 2004; Chambers and Yi, 1999; Träino et al., 2004; Wang et al., 1998; Wasserman et al., 1997) because this surface is stoichiometric and non-polar.

The surface was known to hydroxylate at water vapour pressures greater than 10⁻⁴ Torr. It was believed that the surface was fully hydroxylated in aqueous conditions (Liu et al., 1998). We noticed that there several different hydrated hematite (0 0 0 1) surfaces are known (Mason et al., 2009; Souvi et al., 2013; Wasserman et al., 1997), and in the consideration of our model and the ambient environment, the surface model was built symmetrically so that there is no net surface dipole. The construction of hydrated hematite (0 0 0 1) surface is presented detailedly in Support Information, Text S2 and Fig. S1. After geometry optimization, we obtained a surface relaxed structure that was similar to the structure found in a previous study (Kerisit, 2011; Mason et al., 2009).

The next consideration is the models of the PFSA and hydrated hematite (0 0 0 1) surface in environment for MD simulation. The calculated pKa of PFOS is ~3.27, and 3M confirm that the substance is considered to be a strong acid (Brooke et al., 2004). The reported pKa values of PFBS and PFBES are all 0.14 (Steinle-Darling and Reinhard, 2008), and a previous study has suggested that the pKa values for PFASs are below 0.3 (Vierke et al., 2013). Under the ground water pH, considering the Henderson-Hasselbalch equation:
pH = pKa + log ([A⁻]/[HA])

[A⁻]/[HA] = 10^{pH-pKa}

It is obvious that PFSAs in the water are dissociated. The dissociated state and the non-dissociated state will give rise to different interactions with water and the interface, leading to different results obtained by computational methods, and we used the ionic state to be a close representation of the real environmental conditions.

PFSAs in the environment (e.g., soil, sediment, water) are always surrounded by water molecules. Therefore, it is reasonable to assume that any “realistic” model should include water molecules. The water molecules will present a regular arrangement at the hematite/water interface, and the atomic-level structure of water at surface is an important factor for controlling the interfacial reactions (Kenisit, 2011). The degree of ordering decreases with increasing distance from the surfaces and the ordered water extends at least 1 nm into the bulk fluid on the (0 0 0 1) hematite surface (Catalano, 2011). The existence of PFSAs also influences the distribution of water (Support Information, Text S3 and Fig. S2). Based on the above, we replicated the (1 x 1) two-surface hydrated cell into a (5 x 5) supercell with a vacuum layer of 60 Å filled by liquid with a density of 1.01 g/cm³ that consists of water molecules, one PFSAs ion and a potassium ion. The potassium ion was added to equilibrate the charge. This resulted in a large cell of 25.1445 x 25.1445 x 77.6730 Å³ that contained more than a thousand water molecules. An example of the PFSAs-containing system is shown in Fig. 1.

Adsorption experiments were carried out to support the calculation results. The experimental methods are described in Support Information, Text S4.

3. Results and discussion

3.1. Adsorption configuration

We present the results for PFSA adsorption on the hydrated hematite (0 0 0 1) surface obtained by MD simulations. During the simulations, PFSAs ion apparently first move randomly in the aqueous solution. The configurations and orientations of the PFSAs molecules were not constant or regular, but in general, they did show some trends of motion in the Z direction toward the surface. PFBS, PFPS, PFHxS, PFHpS and PFOS adsorbed after a simulation time of 1.7 ns, 2.38 ns, 3.76 ns, 1.2 ns and 1.64 ns, respectively.

Fig. 2 shows the snapshots of the adsorbed frames that most configurations of the PFSAs ions in dynamic simulations lying on the hydrated hematite (0 0 0 1) surface. The adsorption is not constant. We noted that in the MD simulation, all PFSAs were flexible, could lie, twist or float on the hematite surface and did not stay in a fixed location. We speculated that the fluorine in the PFSAs tail (perfluorooalkyl chain) plays an important role in the system. The hematite surface is highly hydrated resulting in similar surface morphology in the entire surface region. Therefore, at 298 K, PFSAs has the opportunity to transform and move from between the nearby regions.

It is interesting to see that the PFSAs adsorption configurations are not similar to those of alkyl sulfonic acid (e.g., sodium dodecyl sulfate SDS) adsorption on the hematite surface. The adsorption conformation of SDS likely has a different geometry with the first sorbate layer orienting the hydrophobic tail toward the solution (Gao and Chorover, 2010). A previous study focusing on the adsorption of sodium laurate on hematite particles showed that the laurate “stands” on the surface with the head toward the hematite and tail toward the aqueous solution (Chernysylova et al., 2011). As shown in Fig. 2, in our simulations, most of the adsorbed PFSAs configurations are not “vertical”. They interact with surface in the “horizontal” manner. The tail of both short chain PFSAs and PFOS is attached to the hydrated hematite surface with several hydrogen bonds of F···H. This result is similar to the previous study that showed the adsorption isotherms for PFOS onto Ottawa sand, and demonstrated that the long axis is parallel to the surface on the basis of the molecule size and the level of coverage (Johnson et al., 2007). Zhao et al. (2014) concluded that PFOS has a longer carbon chain and could contact with more active sites of humic substances compared with PFHxS, indicating that the carbon chain plays an important role in adsorption. There were also a few configurations where the head group was oriented to the surface with a carbon chain bridge connected to the attached tail, similar to the “U”-shape (the configuration is not shown). Moreover, while a few trajectories indicated that the “stand” adsorption model did exist, it desorbed or changed configuration after a few picoseconds, demonstrating that the interaction between the PFSAs head group and the surface is not sufficient to hold PFSAs to stand stably.

3.2. Adsorption stability

As a preliminary analysis of the adsorption models, we noticed that the PFSAs adsorption structures on the hydrated hematite surface showed some unique styles. To verify whether the adsorption structures are stable, we performed another 1 ns dynamic simulations for each system. Fig. 3 shows the relative concentration of fluorine atoms during the dynamic simulations, where the top of the hydrated hematite (0 0 0 1) surface was set as zero point. Data were collected in the specified direction vertical to the surface. As shown in Fig. 3, there is a sharp peak of fluorine atom in PFSAs tail located around 1.8 Å from the surface, indicating that PFSAs adsorb onto the surface. And there is another peak.
around 4 Å, because of the PFSA's tail structure that have fluorine atoms on the other side. The fluorine atom cannot be detected at the distance above 10 Å. The relative concentration of fluorine atoms suggest that all the PFSA in this study adsorb on the hydrated hematite surface strongly.

3.3. Hydrogen bond

As the relative concentration peak shown in Fig. 3, it is natural to consider the interaction between fluorine atoms and hydrated surface caused by hydrogen bonds (HB). Subsequently, we identify the hydrogen bond formed between the F atoms of PFSA and the O atoms of the hydrated hematite surface using the following geometric criteria to determine whether an HB was formed: \( d_F \ldots \alpha < 3.5 \text{ Å} \) and \( 120^\circ \leq \theta_{\text{O-H-F}} \leq 180^\circ \) (Lu et al., 2010). The HBs from 400 stable adsorbed frames during 1 ns of the simulations were counted. Fig. 4 shows the distribution of the number of HBs for different length ranges of the PFSA. These observed results clearly show a trend of increasing total number of HBs with increased carbon chain length. Generally, the values of the total number of HBs evaluated using different HB lengths also exhibit this tendency. For PFSA, it is easy to figure out that because longer carbon chains have greater number of fluorine atoms, these chains have more opportunities to form HBs between the F atoms of PFSA and the O atoms of the hydrated hematite surface. At the same time, the adsorption strength can be enhanced by increasing number of HBs. Moreover, the distribution of strong hydrogen bond (simply classified by the distance smaller than 2.5 Å) also makes a difference. Among the studied PFSA, the PFOS has the longest carbon chain with the most F atoms to form hydrogen bond, and the short chain PFSA have much shorter carbon chains with fewer F atoms, leading to a relatively weaker adsorption, which is consistent with our experimental result in Support Information Text S4.

3.4. DFT calculations of adsorption

In this section we construct a system without water molecules to decrease the computational cost and to only consider the interaction between the hydrated hematite (0 0 0 1) surface and PFSA. The computational parameters were set to lower accuracy to perform the calculations in suitable time. The modified parameters are as follows: the integration over the Brillouin zone was performed with a \( 1 \times 1 \times 1 \) k point mesh generated following the Monkhorst-Pack scheme. The calculations were performed within the generalized gradient approximation (GGA) proposed by Perdew, Burke and Ernzerhof (PBE). In the calculation of hematite surface and PFSA, other changed details were electronic energy SCF tolerance, \( 1.0 \times 10^{-6} \) eV/atom; energy change per ion, \( \Delta E/\text{ion} \), \( 1.0 \times 10^{-5} \) eV/atom; maximum force, 0.03 eV/Å; and maximum displacement, \( 1.0 \times 10^{-3} \) Å.

Six different adsorption configurations are shown in Fig. 5. Three PFSA including PBBS, PFHxs and PFOS were chosen for typical case studies. The configurations A, B, C were constructed by MD simulations of hematite surfaces and PFSA without the water molecules. Configurations D, E, F were some of the stable adsorption configurations obtained from previous MD simulations by simply deleting the water molecules. We note that there are many other possible adsorption configurations of PFSA on the hematite surfaces, and the presented two configurations of each PFSA only represent the typical possible adsorption behaviour of PFSA. Moreover, under the ambient environment conditions water molecules do exist; nevertheless, to reduce the computational and time cost, and for the purpose of exploring the interaction between the hematite surface and PFSA, the water molecules were ignored.

The adsorption energy (\( E_{\text{ads}} \)) is derived according to the
following equation:

$$E_{\text{ads}} = E_{\text{surface after adsorption}} - (E_{\text{surface}} + E_{\text{adsorbate}})$$

where $E_{\text{surface after adsorption}}$, $E_{\text{surface}}$, and $E_{\text{adsorbate}}$ represent the total energy of the system, the energy of the hydrated hematite surface, and the energy of the PFSAs, respectively. The calculated energy of each contribution is listed in Table 1. When PFSAs adsorb at the surface without the water molecules, the adsorption energy is negative in both adsorption models, meaning that the adsorption of PFSAs on hydrated hematite reduces the total energy. This result indicates that the PFSAs are able to lie on the surface, and the hydrogen bonds between F and H can lower the system energy. Furthermore, the different adsorption models show a trend that the $E_{\text{ads}}$ of A, B, C are all lower than those of D, E, F, respectively; this may be the result of different adsorption geometries. We note that these configurations may not be the global minimum so it is not appropriate to use these results to assess the best structure for adsorption in vacuum. In general, both models will lead to a drop in the total energy. And certainly, the adsorption model with the long axis of PFSAs parallel to the surface can occur. Moreover, in aqueous conditions, fluorine atoms interact with surrounding water molecules as well as the hydrated surface. We roughly calculated the interaction energy between two water molecules and the interaction energy between a water molecule and a PFBS molecule. We

| Adsorption type | $E_{\text{adsorbate}}$ (eV) | $E_{\text{surface}}$ (eV) | $E_{\text{surface after adsorption}}$ (eV) | $E_{\text{ads}}$ (eV) |
|----------------|----------------------------|--------------------------|-----------------------------------------------|-------------------|
| A              | -8179.568                  | -189436.254              | -197616.860                                   | -1.038            |
| B              | -11136.985                 | -189436.254              | -200574.018                                   | -0.779            |
| C              | -14094.373                 | -189436.254              | -203531.394                                   | -0.767            |
| D              | -8179.568                  | -189436.254              | -197616.315                                   | -0.493            |
| E              | -11136.985                 | -189436.254              | -200573.815                                   | -0.576            |
| F              | -14094.373                 | -189436.254              | -203530.826                                   | -0.199            |

* Adsorption types are shown in Fig. 5.
found the optimized distance between the hydrogen atom in the water molecule to the oxygen atom in the other water molecule was approximately 2 Å, and the reduction in the energy was 0.21 eV–0.27 eV. The optimized distance between the hydrogen atom in the water molecule to the F atom on PFASs was in the 2.5 A-2.8 Å range, and the reduction in the energy was 0.009 eV–0.014 eV, an order of magnitude lower than that the interaction between two water molecules. Therefore, the water molecules would tend to aggregate and exclude the PFASs. Moreover, the adsorption energy of PFASs on hematite overwhelmed the interaction between PFASs and water molecules.

The electron density configuration gives us an intuitive picture of the PFASs adsorption on the hematite surface. Fig. 6 shows the three-dimensional isosurface plot of PFOS on the surface as an example with the isovalue of 0.05 e/Å³. The electron cloud of the F atom overlaps parts of the H atom on the hydrated hematite (0 0 0 1) surface, indicating that F and H atoms play an important role during the adsorption process explicitly, also indicating that the hydrogen bond interaction participates in the adsorption.

Analysis of the electron structure based on the density of states (DOS) elucidates the nature of the bonds between the adsorbed PFASs and the hydrated hematite (0 0 0 1) surface. Fig. 7 shows the partial density of states (PDOS) of the F atom on the tail of PFHxS as an example. Comparing the F atom before and after the adsorption, it can be learned from the PDOS that in both adsorption type B and type E, all the electronic density of states are lowered, with the peaks appearing between ~15 eV and ~5 eV and ~30 eV to ~26 eV. This indicates the energies of the adsorbed F atoms are considerably lowered. Fig. 8 shows the PDOS of F atom on PFHxS tail in adsorption type B and the adjacent H atoms on the surface. The p orbital of F atom is mainly localized at ~11 eV to ~9 eV and ~8 eV to ~5 eV, where the s orbital of H atoms also have distribution. The overlap of the PDOS is one piece of evidence for bond formation between atom pairs (Guo et al., 2011), and hydrogen bonds formed by the overlap of F (p) orbital and H (s) orbital are sorbent-sorbate interactions which reveal the mechanism of PFASs adsorbed on hydrated hematite surface.

3.5. Discussion

In the present study, we simulated a situation that was similar to the real environment to infer the PFASs adsorption behaviour. The PFASs adsorption model showed that the long axes lay along the surface, which is dissimilar to the adsorption geometries of alkyl sulfonic acid (e.g., sodium dodecyl sulfate SDS) and laurate on hematite surface (Chernyshova et al., 2011; Gao and Chorover, 2010). Johnson et al. (2007) speculated that the PFOS adsorbs onto Ottawa sand with the long axis parallel to the surface for the monolayer coverage; this is identical to our model. Considering the size of PFOS and assuming that there is no space between the molecules, a density of 4 PFOS/nm² should be obtained on the surface. In our simulation system, the density is only 0.16 PFOS/nm² on the hematite surface, far below the coverage, and the adsorption model is consistent with the previous experimental study.

In recent years, a few studies reported that the short chain PFASs are more mobile than PFOS in the environment. Based on the organic carbon normalized distribution coefficients (KOC), Zhou et al. (2013) speculated that the short chain PFAAs have lower adsorption potentials than PFOS and PFOA. Vierke et al. (2014) investigated the transport of PFAAs using a water-saturated sediment column and found higher mobilities for shorter chain PFAAs and an increasing trend of partition coefficients with increasing chain length. The analysis of our MD simulations was in good agreement with these experimental results.

The negative adsorption energies remind us that the configurations are energy permitted, but the obtained values need to be compared carefully for the adsorption configuration and consideration of different conditions. We also observed that the interface between water and the solid phase is complicated. In addition to hematite, the solid phase such as sediment and sludge plays an important role in the transport and fate of organic pollutants, and the mechanism needs to be investigated further.

4. Conclusions

In the present study, we used high computational methods to construct a system based on a hydrated hematite (0 0 0 1) surface as a model system to investigate the adsorption of PFASs that occurred in environmental conditions. The molecular dynamic simulations showed that all PFASs share the same adsorption configuration as the long carbon chains parallel to the surface. The formation of hydrogen bonds between F atom on PFASs and H atoms on hydrated hematite surface helps to stabilize the unique configuration. Nevertheless, the water molecules trend to form hydrogen bonds between themselves. As a result, the sorption capacity increases with the increasing C–F chain length. Moreover, the adsorption energy and partial density of states (PDOS) analysis demonstrate the adsorption mechanism in between physSa and chemical adsorption because the hydrogen bonds formed by the overlap of F (p) orbital and H (s) orbital are weak intermolecular interactions while the physical adsorption are mainly ascribed to the electrostatic interactions.

Nowadays, these short chain PFASs have given significant contributions to the total PFASs in environment. The data of human and biota exposure to these chemicals are still limited, and considering the persistent characteristic accompanies with the transfer ability of short chain PFASs, it is therefore irresponsible to release them into the environment today, and potential problems of short chain PFASs should be brought to the forefront.

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Appendix A. Supplementary data

Supplementary data related to this article can be found at http://dx.doi.org/10.1016/j.watres.2017.02.024.

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