Surrounding Greenness and Exposure to Air Pollution During Pregnancy: An Analysis of Personal Monitoring Data

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BACKGROUND: Green spaces are reported to improve health status, including beneficial effects on pregnancy outcomes. Despite the suggestions of air pollution–related health benefits of green spaces, there is no available evidence on the impact of greenness on personal exposure to air pollution.

OBJECTIVES: We investigated the association between surrounding greenness and personal exposure to air pollution among pregnant women and to explore the potential mechanisms, if any, behind this association.

METHODS: In total, 65 rounds of sampling were carried out for 54 pregnant women who resided in Barcelona during 2008–2009. Each round consisted of a 2-day measurement of particulate matter with aerodynamic diameter ≤ 2.5 μm (PM2.5) and a 1-week measurement of nitric oxides collected simultaneously at both the personal and microenvironmental levels. The study participants were also asked to fill out a time–microenvironment–activity diary during the sampling period. We used satellite retrievals to determine the surrounding greenness as the average of Normalized Difference Vegetation Index (NDVI) in a buffer of 100 m around each maternal residential address. We estimated the impact of surrounding greenness on personal exposure levels, home-outdoor and home-indoor pollutant levels, and maternal time-activity.

RESULTS: Higher residential surrounding greenness was associated with lower personal, home-indoor, and home-outdoor PM2.5 levels, and more time spent at home-outdoor.

CONCLUSIONS: We found lower levels of personal exposure to air pollution among pregnant women residing in greener areas. This finding may be partly explained by lower home-indoor pollutant levels and more time spent in less polluted home-outdoor environment by pregnant women in greener areas.

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About half of the global population now lives in cities and it is estimated that by 2030 three of every five persons will live in urban areas (Fuller and Gaston 2009; Smith and Guarinzo 2009). Urbanization has led more people to live in environments that are generally more polluted and less green (Cohen et al. 2005; Fuller and Gaston 2009; Grimm et al. 2008; Tzoulas et al. 2007). Urban air pollution is not only linked with increased mortality and morbidity in adults and children (Cohen et al. 2005) but also has been demonstrated to adversely affect fetal development (Shah et al. 2011; Vrijheid et al. 2011). For example, maternal exposure to ambient air pollution has been associated with the risk of low birth weight, preterm birth, intrauterine growth retardation, and congenital anomalies (Shah et al. 2011; Vrijheid et al. 2011). On the other hand, green spaces have been reported to improve both perceived and objective physical and mental health and well-being (Bowler et al. 2010; Maas 2008). More recently, studies have suggested that these spaces have some beneficial effects on pregnancy outcomes (Dadvand et al. 2011; Donovan et al. 2011).

Specifically, these studies, conducted in the United States and Europe, reported that higher surrounding greenness of maternal residential address was positively associated with an increase in birth weight (Dadvand et al. 2011; Donovan et al. 2011).

Although, the underlying pathways of the effects of green spaces on health are not fully understood, increased physical activity, increased social contacts, reduced psychophysiological stress and depression, decreased noise, microclimate regulation (i.e. moderation of ambient temperature and urban heat island effects), and reduced air pollution levels have been suggested to be involved (Bowler et al. 2010; Gill et al. 2007; Greenspace Scotland 2008; Health Council of the Netherlands 2004; Lee and Maheswaran 2010; Maas 2008; Maas et al. 2009a, 2009b; Nowak et al. 2006; Paolietti et al. 2011; Su et al. 2009, 2011), the available evidence on the impact of surrounding greenness on personal exposure to air pollution is nonexistent. This impact, if any, is of importance because personal exposure to air pollution is a complex function of microenvironmental pollutant levels and personal behavior, both of which could be associated with surrounding greenness.

The aim of our study was to investigate the association between surrounding greenness and personal exposure to air pollution among pregnant women and to explore the potential mechanisms behind this association, if any.

Materials and Methods

Study population. The study population consisted of 54 Spanish-speaking pregnant women who lived in Barcelona, Spain, from November 2008 to November 2009. The study population

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women were recruited from those attending the obstetrics department of the Hospital Clinic of Barcelona for their first, second, or third pregnancy visits. Hospital Clinic of Barcelona is a major university hospital in Barcelona with a catchment area of about one million inhabitants (Figueras et al. 2008). We randomly selected 16 dates from 2008 to 2009, and pregnant women who had an appointment with the hospital within those dates were contacted by phone and invited to participate in the study (n = 434). Women who were interested in participating in the study were then recruited by the study technician on their next visit to the hospital.

**Air pollution sampling.** Nitric oxides (NO\textsubscript{x}) and particulate matter with aerodynamic diameter \( \leq 2.5 \, \mu m \) (PM\textsubscript{2.5}) were used as indicators of air pollution in our study. Each sampling round consisted of a 2-day measurement of PM\textsubscript{2.5} and a 1-week measurement of NO\textsubscript{x} simultaneously for both personal and microenvironmental levels. For 11 participants, we carried out two sampling rounds.

**Personal exposure levels.** During the first 48-hr of the 1-week sampling period, which always occurred on weekdays, participants were asked to wear a small backpack containing a portable particle monitor when they were awake and to place the monitor next to their bed when they were sleeping. The portable particle pump (BGI400S pump; BGI Incorporated, Waltham, MA, USA) operated continuously during this period, collecting PM\textsubscript{2.5} using a GK2.05 sampler (BGI Incorporated) and a 37-mm Teflon filter (Pall Corporation, East Hills, NY, USA) with a 50% cut-point of 2.5 \( \mu m \) at a flow rate of 4 L/min (Van Rooijen et al. 2006). The flow rate was adjusted at the beginning of each sampling round to 4 L/min with a rotameter (model RM67, BGI Incorporated) and checked at the end of that sampling round to make sure that it had remained at 4 L/min during the course of sampling. In addition, the participants were asked to wear a passive sampler (PS-100; Ogawa & Co. USA, Inc., Pompano Beach, FL, USA) during the entire week to measure personal levels of NO\textsubscript{x}.

**Microenvironmental levels.** Pollutant levels were measured within (home-indoor levels) and outside (home-outdoor levels) the participants’ homes. The home-indoor monitors were placed in the living room, while the home-outdoor monitors were installed outside a window or in the terrace/balcony of maternal homes. Similar to the personal measurements, the microenvironmental PM\textsubscript{2.5} concentrations were measured by a portable particle monitor for the first two days and NO\textsubscript{x} levels were measured by a passive sampler during the whole week of the sampling round at fixed locations.

**Time–activity data.** To characterize the time that the participants spent in various microenvironments and activities, we used a modified version of the EXPOLIS (Air Pollution Exposure Distributions of Adult Urban Populations in Europe) Time–Microenvironment–Activity Diary (TMAD) (Hänninen et al. 2004), which we enhanced to include self-reported levels of physical activity for each activity. Each participant was asked to record the microenvironment–activity category for every 30 min of her sampling period. The microenvironment categories included in the TMAD were walk, bike, motorcycle, car/taxi, bus, tram, metro, and train (“in transfer” categories) and indoors at home, outdoors at home, indoors at work, outdoors at work, indoors at another location, and outdoors at another location (“not in transfer” categories) (Hänninen et al. 2004). The diary also included data on smoking (active or passive), use of a nonelectrical heater, and the application of the kitchen hood while cooking with nonelectrical appliances (see Supplemental Material, Figure S1 [http://dx.doi.org/10.1289/ehp.1104609]).

**Surrounding greenness.** To determine the surrounding greenness, we used the Normalized Difference Vegetation Index (NDVI) derived from the Landsat Enhanced Thematic Mapper Plus (ETM+) data at 30 m \( \times 30 \) m resolution (see Supplemental Material, Figure S2 [http://dx.doi.org/10.1289/ehp.1104609]). The ETM+ Landsat data were acquired for 10 August 2000 covering path 197 and row 031 (i.e., the scope of Barcelona). NDVI is an indicator of greenness based on land surface reflectance of visible (red) and near-infrared parts of spectrum (Weier and Herrig 2011). NDVI ranges between –1 and 1 with higher numbers indicating more greenness. Surrounding greenness was abstracted as the average of NDVI in a buffer of 100 m around each maternal place of residence geoocoded according to the address at the sampling time. The choice of buffer size was based on the findings of two recent studies on the association between green spaces and pregnancy outcomes showing that only immediate surrounding greenness (buffers of 50 m and 100 m) was related with pregnancy outcomes (Dadvand et al. 2011; Donovan et al. 2011). This choice was also informed by previous reports suggesting that the impact of green spaces on reducing air pollution levels mostly occurs within immediate vicinity of these spaces and decays rapidly as the distance from green spaces increases (Givoni 1991).

**Statistical analysis.** We estimated the impact of surrounding greenness on personal exposure levels, home-outdoor and home-indoor pollutant levels, and maternal time-activity. For these associations, we developed linear regression models to estimate the change in the outcome associated with an interquartile range (IQR) increase in surrounding greenness (NDVI). We used the measured values for the outcomes without transformation to produce interpretable results and capture the entire range of the variation. We checked the homoscedasticity and normality of the regression residuals and confirmed the robustness of our regressions to these assumptions (data not shown).

**Personal exposure levels.** For each pollutant, we first modeled personal exposure levels against surrounding greenness using the regionwide average of pollutant levels during the sampling period for each participant as an offset to account for temporal variation in background pollutant levels. To calculate regionwide averages, the daily pollutant levels, which were measured by seven monitoring stations for NO\textsubscript{x} and three monitoring stations for PM\textsubscript{2.5} across Barcelona, were averaged for each day of the study period. Each participant was then assigned the mean of the daily pollutant levels for her sampling period. We further adjusted the estimates for personal exposure for the time spent at home (sum of time spent at home-indoor and home-outdoor), time spent in transfer, smoking (active and passive), use of gas-cooking appliances, and the MEDEA [Construcción de un índice de privación a partir de datos censales en grandes ciudades españolas (Proyecto MEDEA)] index of neighborhood deprivation (Domínguez-Berjón et al. 2008). Maternal socioeconomic status (SES) could be associated with both air pollution levels (personal, indoor, and outdoor) and surrounding greenness. We included indicators of SES at both individual and neighborhood levels in our analyses. For individual SES we adjusted for number of inhabitants at home (an indicator for overcrowding), as well as use of gas-cooking appliances and maternal exposure to tobacco smoke, which might also be indicators of SES. For the neighborhood SES, we used the MEDEA index which measures deprivation at the census tract level based on five domains including percentage of manual workers, temporary workers, people with low education (overall), young population with low education, and unemployment (Domínguez-Berjón et al. 2008). These domains have been shown to explain 75% of the variability of all socioeconomic variables available in the Spanish census (Domínguez-Berjón et al. 2008). In the 2001 Census, there were 1,491 census tracts across the city of Barcelona with a median area of 0.02 km\(^2\) and population of 992.

**Microenvironmental pollutant levels.** Pollutant-specific models were developed using home-outdoor pollutant levels as the outcome; surrounding greenness, surrounding traffic intensity, the height of the monitor, and the MEDEA index of neighborhood deprivation as predictors; and the regionwide average of pollutant levels during the sampling
period for each participant as an offset. Traffic flow in each street was obtained from traffic flow models developed by the Barcelona City Council (Ajuntament de Barcelona 2007). For each participant, the lengths of all streets falling within the buffer of 100 m around her place of residence were abstracted and multiplied by the corresponding flow for each street. The surrounding traffic intensity (vehicle kilometers traveled) was then calculated by summing these values. The Pearson’s correlation coefficient between surrounding greenness and surrounding traffic intensity within the 100 m buffer was \(-0.08\) (\(p = 0.51\)).

For each pollutant, we modeled home-indoor levels against the surrounding greenness, smoking, use of gas-cooking appliances, the number of inhabitants at home, the temperature at home-indoor on the first day of sampling round, and the MEDEA index of neighborhood deprivation, while using region-wide average of pollutant levels during the sampling period for each participant as an offset.

**Table 1. Descriptive statistics of the study participants (n = 54).**

| Variable                        | Median (IQR) |
|---------------------------------|--------------|
| Personal exposure (µg/m³)       | 23.2 (9.9)   |
| NO₂                             | 56.6 (34.1)  |
| Home-indoor levels (µg/m³)      | 20.1 (13.5)  |
| NO                              | 55.1 (40.8)  |
| Home-outdoor levels (µg/m³)     | 17.7 (9.5)   |
| NO₂                             | 49.9 (26.3)  |
| Home surrounding NDVI average   | -0.283 (0.049) |
| 100 m buffer                    | -0.275 (0.031) |
| 250 m buffer                    | -0.267 (0.042) |
| Time spent (min/day)            | 1.9 (15)     |
| Transfer                        | 67.5 (50.8)  |
| Cooking fuel [n (%)]            |              |
| Natural gas                     | 39 (72.2)    |
| Butane gas                      | 7 (13.0)     |
| Electric/hot gas                | 8 (14.8)     |
| Smoking [n (%)]                 |              |
| Yes                             | 24 (44.4)    |
| No                              | 30 (55.6)    |
| Home-indoor temperature (°C)²   | 24.5 (5.0)   |
| Monitor height (m)              | 9 (6)        |
| MEDEA index                     | 0.20 (1.44)  |
| No. of inhabitants at home      | 3 (2)        |

*Data are median (IQR), unless otherwise noted. Cumulative concentration for 2 days of sampling. Cumulative concentration for 1 week of sampling. Measured on the first day of sampling.*

**Sensitivity analyses. Buffer size for abstracting surrounding greenness.** We chose a buffer of 100 m around each maternal residence to calculate surrounding greenness. To evaluate the robustness of our findings to alterations in this buffer size, we abstracted the surrounding greenness in buffers of 250 m and 500 m and repeated all the aforementioned analyses using these alternative exposure variables plus alternative traffic intensity variables for buffers of 250 m and 500 m, respectively. The results were expressed for a change in the IQR of the surrounding greenness of the subjects in each buffer size.

**Surrounding greenness of the working place.** Of the 54 study participants, 18 were working outside of their homes during their sampling period. The working address was known and geocoded for 17 participants and the surrounding greenness (100-m buffer) was extracted for these addresses. We calculated a home–work surrounding greenness index by averaging the surrounding greenness of maternal residential and working addresses weighted by the time that participants spent in each place. We then modeled personal exposure levels against this index together with smoking, use of gas-cooking appliances, time spent in transfer, MEDEA index of neighborhood deprivation, and a binary (yes/no) variable determining whether the participant was working during the sampling period.

**Repeated sampling rounds.** Our analysis was based on 65 observations from 54 participants including two sampling rounds for a subset of 11 participants. We considered a gap of at least 1 month between two sampling rounds for each participant to minimize the effect of repetitive measurements. To evaluate the effect of second sampling rounds on our findings, we carried out a sensitivity analysis including only the first sampling round for each participant in the models for personal, home-indoor, and home-outdoor levels.

**Ethics approval.** Ethics approval (No. 2008/3115/I) was obtained from the Clinical Research Ethical Committee of the Parc de Salut MAR, Barcelona, Spain, to carry out this study. All participants gave their written informed consent prior to the study.

**Results.** Of the 434 women contacted by the study technicians, 54 agreed to participate in the study. A total of 65 air pollution measurements were carried out, with two sampling rounds for 11 participants and a single round for 43 participants. One of the participants (one sample) with metallurgy processing nearby her home was identified as an outlier and excluded from the analyses. One sampling round was performed during the first trimester, 23 during the second trimester, and 41 during the third trimester of participants’ pregnancies. As presented in Table 1, 24 (44.4%) participants reported exposure to tobacco smoke (passive or active) and 46 (85.2%) participants reported use of gas-cooking appliances during the sampling period.

Median home-indoor pollutant levels were generally higher than median home-outdoor levels (Wilcoxon signed-rank test \(p\)-value of 0.06 for PM₁.5 and <0.01 for NO₂ (Table 1). On average, participants spent 73.3% of their time at home–indoors and 1.1% at home–outdoors. Participants working outside of the home spent 14.4% of their time at work (including indoor and outdoor time). The median (minimum and maximum) of surrounding NDVI (100 m buffer) was \(-0.28\) (\(-0.32, -0.14\)) for residential addresses and \(-0.27\) (\(-0.36, -0.21\)) for places of work.

Higher residential surrounding greenness was associated with lower average levels of personal PM₁.5 exposures in both unadjusted and adjusted models (Table 2). The increase in surrounding greenness of maternal residential addresses was also associated with a decrease in the average home-indoor and home-outdoor PM₁.5 levels, with a statistically significant (\(p < 0.05\)) association for home-indoor PM₁.5 and nearly statistically significant association for home-outdoor PM₁.5 (Table 2). For NO₂, none of the associations attained statistical significance, but their directions were consistent with those of PM₁.5 (Table 2).

For maternal time–activity patterns, an interquartile increase in surrounding greenness was associated with a nearly statistically significant 12-min/day increase in time spent home–indoors (95% confidence interval (CI): 0, 24; \(p = 0.07\)) but was not significantly associated with time spent home–indoors (6-min increase, 95% CI: \(-66, 71\); \(p = 0.94\)). Trimester of sampling was not significantly associated with the time spent at home–indoors or home–outdoors (Kruskal–Wallis test \(p\)-values of 0.44 and 0.24, respectively).

The results from the sensitivity analyses evaluating the impact of different buffer sizes on our findings were generally consistent with our main findings; however, for home-outdoor PM₁.5, the association became stronger and attained statistical significance when surrounding greenness was based on a 500-m buffer (Table 2). For personal PM₁.5 exposure, the association seemed to weaken with larger buffer sizes, with a statistically nonsignificant association for a 500-m buffer. For NO₂ associations remained statistically nonsignificant for all buffer sizes (Table 2).

An IQR (0.57) increase in the home–work surrounding greenness index was associated with a 5.4-µg/m³ decrease (95% CI: \(-10.2, -0.6\)) in average personal PM₁.5 levels and a statistically nonsignificant 5.3-µg/m³ decrease (95% CI: \(-17.8, 7.3\)) in average personal NO₂ levels.
The exclusion of data from second sampling rounds for the 11 participants with two sampling rounds did not result in a notable change in our findings (see Supplemental Material, Table S1 [http://dx.doi.org/10.1289/ehp.1104609]).

Discussion

To our knowledge, this is the first study to report on the association between surrounding greenness and measured personal exposure to air pollution. Our study sample consisted of 54 pregnant women who resided in Barcelona from 2008 to 2009. The analysis was based on data obtained from time–activity diaries and air pollution sampling rounds that included personal and microenvironmental measurements of PM$_{2.5}$ (2 days) and NO$_x$ (1 week).

To measure surrounding greenness, we used averages of NDVI within a buffer of 100 m around each maternal residential address. Higher residential surrounding greenness was associated with lower personal home-indoor and home-outdoor PM$_{2.5}$ levels and with more time spent home-outdoor. The results for NO$_x$ were less conclusive.

We observed a reduction in average personal exposure to PM$_{2.5}$ associated with an increase in greenness surrounding the maternal residential addresses. There was also an indication for a similar association between surrounding greenness and average personal exposure to NO$_x$ but the association did not attain statistical significance. NO$_x$ participate in complex photochemical reactions with volatile organic compounds (VOCs) emitted by plants (Kesselmeier and Staudt 1999) and this could make identifying trends for NO$_x$ and green space more challenging.

We are unaware of previous studies on the impact of surrounding greenness on measured personal exposure among pregnant women or in the general population. In a recent analysis of modeled ambient concentrations of pollutants [nitrogen dioxide (NO$_2$), PM$_{2.5}$, and ozone] in and around parks in the Los Angeles metropolitan area, California, Su et al. (2011) found that PM$_{2.5}$ and NO$_2$ were higher in neighborhoods adjacent to the parks. Their finding was likely due to the very particular circumstances of the Los Angeles area where parks are predominantly located near highways with heavy traffic. Our measure is different in that the presence of green space as indicated by the NDVI index is a more continuous measure evenly distributed across the city and is not related to proximity to roadways with heavy traffic. In addition, our multivariate analysis of home-outdoor concentrations accounted for traffic intensity in the immediate surroundings of the homes (100-m buffer). In the Los Angeles study, Su et al. (2011) did find, however, that modeled pollutants were lower inside the parks compared with the region as a whole, which perhaps better reflects the greenery in the immediate surroundings of the residences of our participants as measured by the NDVI.

The findings for home-outdoor PM$_{2.5}$ levels are in line with the available evidence that has shown that green spaces reduce air pollutant levels by filtering air pollutants, improving urban circulation, and reducing ambient temperature (Akbari 2002; Givoni 1991; Nowak et al. 2006; Paoletti et al. 2011; Su et al. 2011).

To our knowledge, there is no previous report quantifying the association between surrounding greenness and home-indoor pollutant levels. Our observed reduction in home-indoor PM$_{2.5}$ levels could have been secondary to decreases in home-outdoor PM$_{2.5}$ levels.

The study participants spent much of their time home-indoors (73.3%) and our observed lower home-indoor pollutant levels for participants with higher degrees of surrounding greenness could partly explain their lower personal exposure levels. The impact of lower levels of indoor pollution in greener areas is supported by our observed strong correlation between personal and home-indoor PM$_{2.5}$ levels (Pearson’s correlation coefficient of 0.78, p < 0.01). Furthermore, higher degrees of surrounding greenness were associated with a nearly significant 86% increase in the time that our participants on average spent home-outdoors (14 min). This increase was in line with previous reports suggesting an increase in outdoor physical activity in relation with higher degrees of surrounding greenness (Ellaway et al. 2005). Because home-outdoor pollutant levels were lower than those of home-indoor levels, this increase in time spent in less polluted home-outdoor environment could result in lower personal exposure levels.

We conducted a range of sensitivity analyses, including testing the effect of the buffer size for calculating surrounding greenness on our investigated associations. Our findings appeared to be robust to buffer size; however, for personal PM$_{2.5}$ exposure the associations seemed to weaken with larger buffer sizes. This pattern is in line with findings of previous reports showing that only immediate surrounding greenness was associated with pregnancy outcomes (Dadvand et al. 2011; Donovan et al. 2011). We also conducted a separate analysis of associations with surrounding greenness at work as well as at home using a weighted average of NDVI at both locations for the subset of participants who worked outside of the home, and the results were consistent with our main findings.

Our study faced some limitations. Indoor greenness (e.g., houseplants) has been suggested to improve indoor air quality (Claudio 2011). We did not have data on indoor

### Table 2. Regression coefficients (95% CIs) of change in personal exposure and microenvironmental pollutant levels (µg/m$^3$) associated with an IQR$^a$ increase in the average NDVI within the buffers of 100 m, 250 m, and 500 m around maternal residential addresses.

| Measurements     | 100-m buffer            | 250-m buffer            | 500-m buffer            |
|------------------|-------------------------|-------------------------|-------------------------|
|                  | Regression coefficient (95% CI) | p-Value               | Regression coefficient (95% CI) | p-Value               | Regression coefficient (95% CI) | p-Value               |
| Personal (unadjusted) |                          |                         |                         |                          |                          |                         |
| PM$_{2.5}$       | -5.2 (–9.4, –0.9)        | 0.02                    | -2.4 (–5.0, 0.1)        | 0.06                    | -2.8 (–5.0, 0.3)         | 0.08                    |
| NO$_x$           | -2.6 (–15.3, 10.1)       | 0.89                    | -2.3 (–7.5, 3.1)        | 0.54                    | -3.2 (–12.4, 6.0)        | 0.49                    |
| Personal (adjusted)$^b$ |                          |                         |                         |                          |                          |                         |
| PM$_{2.5}$       | -5.9 (–10.0, –1.8)       | < 0.01                  | -2.4 (–4.8, 0.0)        | 0.05                    | -2.3 (–5.1, 0.5)         | 0.11                    |
| NO$_x$           | -5.1 (–18.6, 8.4)        | 0.45                    | -3.0 (–10.7, 4.6)       | 0.43                    | -3.6 (–12.9, 5.7)        | 0.44                    |
| Home-indoor$^c$  |                          |                         |                         |                          |                          |                         |
| PM$_{2.5}$       | -6.1 (–10.6, –1.6)       | < 0.01                  | -1.9 (–4.6, 0.8)        | 0.17                    | -2.3 (–5.5, 0.9)         | 0.15                    |
| NO$_x$           | -9.5 (–24.4, 5.3)        | 0.20                    | -4.5 (–13.3, 4.2)       | 0.31                    | -6.7 (–17.3, 3.9)        | 0.21                    |
| Home-outdoor$^d$ |                          |                         |                         |                          |                          |                         |
| PM$_{2.5}$       | -4.4 (–9.5, 0.7)         | 0.08                    | -3.2 (–6.6, 0.2)        | 0.07                    | -5.5 (–10.5, –0.4)       | 0.04                    |
| NO$_x$           | -5.8 (–17.6, 6.0)        | 0.33                    | -5.3 (–14.0, 3.4)       | 0.23                    | -5.6 (–19.5, 8.3)        | 0.43                    |

$^a$0.049 for 100 m buffer, 0.031 for 250 m buffer, and 0.042 for 500 m buffer. $^b$Adjusted for the time spent at home (sum of time spent at home-indoor and home-outdoor), smoking (active and passive), use of gas-cooking appliances, time spent in transfer, and MEDEA index of neighborhood deprivation. $^c$Adjusted for the temperature at home-indoors on the first day of sampling round, the use of gas-cooking appliances, smoking (active and passive), the number of inhabitants, and MEDEA index of neighborhood deprivation. $^d$Adjusted for the traffic intensity in the buffer of 100 m around maternal residential address, the height of the monitor, and MEDEA index of neighborhood deprivation.
greenness and were not able to address it in our analyses. This might confound or modify our estimated associations. Furthermore, our study used remote-sensing–derived NDVI to measure surrounding greenness. Application of this objective measure of greenness enabled our study to take account of small-scale green spaces (e.g., home gardens, street trees, and green verges) in a standardized way; however, NDVI does not distinguish between different types of vegetation. This distinction may be important because there is some evidence that the effect of green spaces on air pollutants is vegetation-dependent, with trees being the most effective and grasses being the least effective (Givoni 1991). The inability of NDVI to distinguish between different vegetation types may have been a source of exposure misclassification in our study. The NDVI map used in this study was based on remote sensing data collected in 2000 while our study was carried out during 2008–2009. To evaluate the temporal validity of our NDVI map, we applied an ecologic map of Barcelona for the year 2004 (Centre for Ecological Research and Forestry Application Application 2004) to abstract the percentage of the green areas over grids of 100 m × 100 m across the Barcelona. We also abstracted the averages of NDVI over the same grids. The strong correlation (Pearson’s correlation coefficient of 0.81, p < 0.001) between these two measures could suggest a satisfactory temporal validity of NDVI map to estimate the surrounding greenness for our study.

Our classification of the time spent by the participants in each microenvironment was based on the TMAD records. Although the participants were asked to provide a TMAD record every 30 min, which could minimize the problem of recall, there might still be some misclassifications if the women filled out their TMAD after longer periods of time (e.g., if they left their TMAD at home when they went to work). We did not equip our participants with global positioning systems (GPSs), so we were not able to validate TMAD records on time spent in microenvironments. This would be helpful in future studies.

Conclusions

We investigated the association between surrounding greenness of the place of residence and personal exposure to PM_{2.5} and NOx among a sample of 54 pregnant women in Barcelona (2008–2009) and observed lower average PM_{2.5} exposure levels associated with higher surrounding greenness. This finding could be partly explained by lower home–indoor pollutant levels (where participants spent most of their time) and more time spent in less polluted home–outdoor environments by the participants in greener areas.

The time–activity patterns of our study sample are not necessarily generalizable to the general population, especially since most of the samplings were carried out during the third trimester of pregnancy when the mothers would be less mobile in comparison with the general population. Because the time–activity pattern plays an important role in personal exposure to air pollutants, the extrapolation of our findings to the general population should be considered with caution. Nonetheless, our findings are suggestive for a possible beneficial effect of green spaces in reducing exposure to air pollution among the general population (in particular for less mobile groups such as the elderly and children) and would therefore encourage further investigation of this effect. This impact would also be of interest for policy makers in planning sustainable development of urban environments. We recommend future studies to be based on larger samples from general populations and rely on better characterization of surrounding greenness (i.e., vegetation type) and take account of indoor greenness.

References

Ajuntament de Barcelona. 2007. TransCAD 2007, Planificació i estudis de mobilitat. Barcelona, Spain:Ajuntament de Barcelona.
Akbari H. 2002. Shade trees reduce building energy use and CO_{2} emissions from power plants. Environ Pollut 116(suppl 1):S119–S126.
Bowler DE, Buyung-Ali LM, Knight TM, Pullin AS. 2010. A systematic review of evidence for the added benefits to health of exposure to natural environments. BMC Public Health 10:456; doi:10.1186/1471-2458-10-456 [Online 4 August 2010].
Centre for Ecological Research and Forestry Application Application 2004. Ecological Map of Barcelona. 3rd ed. Available: http://www.creeat.uab.es/eng/projects/22_272.htm [accessed 12 January 2012].
Claudio L. 2011. Planting healthier indoor air. Environ Health Perspect 119:A426–A447.
Cohen AJ, Anderson HR, Ostro B, Pandey KD, Krzyzanowski M, Künzli N, et al. 2005. The global burden of disease due to outdoor air pollution. J Toxicol Environ Health A 68(13–14):1301–1307.
Dadvand P, de Nazelle A, Figueras F, Basagaña X, Su J, Amoly E, et al. 2011. Green space, health inequality, and pregnancy. Environ Int 40:110–115.
Dominguez-Berjón MF, Borrell C, Cano-Serrall G, Esnaola S, Nolasco A, Passari M, et al. 2008. Construcción de un índice de privación a partir de datos censales en grandes ciudades españolas (Proyecto MEDEA). Gac Sanit 22(3):179–187.
Donovan GH, Michael YL, Butty DT, Sullivan AD, Chase JM. 2011. Urban trees and the risk of poor birth outcomes. Health Place 17(1):390–393.
Ellaway A, MacIntyre S, Bonnefoy X. 2005. Graffiti, greenery, and obesity in adults: secondary analysis of European cross-sectional survey. BMJ 331(7517):511–512.
Figueras F, Meler E, Iraola S, Eixarch E, Coll O, Figueras J, et al. 2008. Customized birthweight standards for a Spanish population. Eur J Obstet Gynecol Reprod Biol 136(1):20–24.
Fuller RA, Gaskon KJ. 2009. The scaling of green space coverage in European cities. Biol Lett 5(3):352–355.
Gill SE, Handlef JF, Ennos AR, Pauleit S. 2007. Adapting cities for climate change: the role of the green infrastructure. Built Environ 33(1):115–133.
Givoni B. 1991. Impact of planted areas on urban environmental quality: a review. Atmos Environ Part B Urban Atmos 25(3):289–299.
Greenspace Scotland. 2008. Health Impact Assessment. Health Impact of Greenspace—A Guide. Available: http://www.greenspacescotland.org.uk/health-impact-assessment. [accessed 27 July 2012].
Griffin NR, Faeth SH, Golubiewski NE, Redman CL, Wu J, Bai X, et al. 2008. Global change and the ecology of cities. Science 319(5864):756–760.
Hänninen GO, Alm SR, Katsumayumi K, Künzli N, Maroni M, Newnhamhuyen MJ, et al. 2004. The EXPOSUL study: implications for exposure research and environmental policy in Europe. J Expo Anal Epidemiol 14(6):440–456.
Health Council of the Netherlands. 2004. Nature and Health: The Influence of Nature on Social, Psychological and Physical Well-Being. The Hague:Health Council of the Netherlands.
Kesselmer J, Staubli M. 1999. Biogenic volatile organic compounds (VOCs): an overview on emission, physiology, and ecology. J Atmos Chem 33:23–88.
Lee AC, Mahesarwan R. 2010. The health benefits of urban green spaces: a review of the evidence. J Public Health 80(3):212–222.
Maas J. 2008. Vitamin G: Green Environments—Healthy Environments. Utrecht, the Netherlands:Netherlands Institute for Health Services Research.
Maas J, Van Dillen SME, Verheij RA, Groenewegen PP. 2008a. Social contacts as a possible mechanism behind the relation between green space and health. Health Place 15(2):586–595.
Maas J, Verheij RA, de Vries S, Spreeuwenberg P, Schellieis FG, Groenewegen PP. 2009b. Morbidity is related to a green living environment. J Epidemiol Community Health 63(12):967–973.
Novellie D, Crane DE, Stevens JC. 2006. Air pollution removal by urban trees and shrubs in the United States. Urban Forestry Urban Greening 4:115–123.
Paoletti E, Bardelli T, Giovannini G, Pecchioli L. 2011. Air quality impact of an urban park over time. Procedia Environ Sci 12:498–516.
Shah PS, Balkhair T, Knowledge Synthesis Group on behalf of the Determinants of Preterm/LBW Births. 2011. Air pollution and birth outcomes: a systematic review. Environ Int 37(2):498–516.
Smith MP, Guarnizo LE. 2010. The health benefits of urban green spaces: a review of the evidence. BMC Public Health 10:456; doi:10.1186/1471-2458-10-456 [Online 4 August 2010].
Su JG, Jarrett M, Beckerman B, Wilhelm M, Ghosh JK, Ritz B. 2009. Predicting traffic-related air pollution in Los Angeles using a distance decay regression selection strategy. Environ Res 109(8):897–906.
Su JG, Jarrett M, de Nazelle A, Wolch J. 2011. Does exposure to air pollution in urban parks have socioeconomic, racial, or ethnic gradients? Environ Res 113(1):319–328.
Tzoulas K, Korpela K, Venn S, Yi-Pelkonen V, Kazmierzak A, Niemela J, et al. 2007. Promoting ecosystem and human health in urban areas using Green Infrastructure: A literature review. Landsc Urban Plan 81(3):167–178.
Van Roosbroeck S, Wichmann J, Janssen NA, Hoek G, van Wijnen JH, Lebrer E, et al. 2006. Long-term personal exposure to traffic-related air pollution among school children, a validation study. Sci Total Environ 368(1–3):585–593.
Vrijheid M, Martinez D, Manzarones S, Dadvand P, Schemnari A, Rankin J, et al. 2011. Ambient air pollution and risk of congenital anomalies: a systematic review and meta-analysis. Environmental Health Perspect 119:598–606.
Weier J, Herring D. 2011. Measuring Vegetation (NDVI & EVI). Available: http://earthobservatory.nasa.gov/Features/MeasuringVegetation/[accessed 12 January 2012].
Whitford V, Ennos AR, Handley JF. 2001. “City form and natural processes”—indicators for the ecological performance of urban areas and their application to Merseyside, UK. Landsc Urban Plan 57(2):91–103.