Diurnal and Seasonal Variations in the Net Ecosystem CO\textsubscript{2} Exchange of a Pasture in the Three-River Source Region of the Qinghai-Tibetan Plateau

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Abstract

Carbon dioxide (CO\textsubscript{2}) exchange between the atmosphere and grassland ecosystems is very important for the global carbon balance. To assess the CO\textsubscript{2} flux and its relationship to environmental factors, the eddy covariance method was used to evaluate the diurnal cycle and seasonal pattern of the net ecosystem CO\textsubscript{2} exchange (NEE) of a cultivated pasture in the Three-River Source Region (TRSR) on the Qinghai-Tibetan Plateau from January 1 to December 31, 2008. The diurnal variations in the NEE and ecosystem respiration (R\textsubscript{e}) during the growing season exhibited single-peak patterns, the maximum and minimum CO\textsubscript{2} uptake observed during the noon hours and night; and the maximum and minimum R\textsubscript{e} took place in the afternoon and early morning, respectively. The minimum hourly NEE rate and the maximum hourly R\textsubscript{e} rate were −7.89 and 5.03 μmol CO\textsubscript{2} m\textsuperscript{−2} s\textsuperscript{−1}, respectively. The NEE and R\textsubscript{e} showed clear seasonal variations, with lower values in winter and higher values in the peak growth period. The highest daily values for C uptake and R\textsubscript{e} were observed on August 12 (−2.91 g C m\textsuperscript{−2} d\textsuperscript{−1}) and July 28 (5.04 g C m\textsuperscript{−2} day\textsuperscript{−1}), respectively. The annual total NEE and R\textsubscript{e} were −140.01 and 403.57 g C m\textsuperscript{−2} year\textsuperscript{−1}, respectively. The apparent quantum yield (α) was −0.0275 μmol μmol\textsuperscript{−1} for the entire growing period, and the α values for the pasture’s light response curve varied with the leaf area index (LAI), air temperature (T\textsubscript{a}), soil water content (SWC) and vapor pressure deficit (VPD). Piecewise regression results indicated that the optimum T\textsubscript{a} and VPD for the daytime NEE were 14.1°C and 0.65 kPa, respectively. The daytime NEE decreased with increasing SWC, and the temperature sensitivity of respiration (Q\textsubscript{10}) was 3.0 during the growing season, which was controlled by the SWC conditions. Path analysis suggested that the soil temperature at a depth of 5 cm (T\textsubscript{soil}) was the most important environmental factor affecting daily variations in NEE during the growing season, and the photosynthetic photon flux density (PPFD) was the major limiting factor for this cultivated pasture.

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Introduction
Grassland ecosystems occupy approximately one-third of the total global land area and form an important component of the earth’s carbon circulation [1]. During the past few decades, ecologists have studied the effects of environmental factors (such as radiation, temperature, water and soil nutrition), biological factors and management measures on the carbon exchange between the land surface and the atmosphere of the grassland ecosystem by using eddy covariance [2, 3], and these ecologists have noted the significance of human activity on the carbon exchange process [4, 5]. The grassland of China occupies approximately 40% of the nation’s total land area and plays an extremely important role in the regional circulation of carbon [6]. However, because the study of China’s grassland carbon flux began late, these studies have mainly focused on the low-lying regions of China [7].

The Qinghai–Tibetan Plateau has drawn considerable attention as the “initiation zone” and the “sensitivity zone” for China’s weather changes [8, 9]. Although there have been reports on the process of carbon exchange between the land surface and the atmosphere and on the carbon exchange mechanisms of the primary natural vegetation types (e.g., alpine meadows and alpine shrubs) over the last several years [10, 11], there have only been a few reports on the carbon exchange process, the source/sink function of planted vegetation (e.g., cultivated grassland) and the mechanisms controlling the exchange among environmental and biological factors.

The Three-River Source Region (TRSR, i.e., the source of the Yangtze, Yellow and Mekong Rivers and well known as the “water tower of Asia”) is located in the hinterland of the Qinghai–Tibetan Plateau. In recent years, the grassland in this region has severely degraded. Statistics indicate that the area that is experiencing moderate and severe degradation has already reached 5.7×10⁶ hm², occupying 55.40% of the total usable grassland area in this region [12]. This degradation can reduce vegetation biomass [13], soil microorganism activity [14] and soil carbon and nitrogen pools [13, 15] and can increase carbon dioxide (CO₂) emissions [16]. It is estimated that during the last 30 years, approximately 1.01 Pg of soil carbon was emitted from the grasslands of the plateau due to changes in land use and grassland degradation [17]. Thus, grassland degradation on the Qinghai–Tibetan Plateau may have an important impact on the carbon balance at both the regional and global scales. To restore grassland, China’s largest demonstration area for “returning grazing land to grassland” was established in the TRSR. By 2005, the cultivated pasture area in the TRSR had already reached 160,000 km² [18]. An increase in cultivated pasture may slow degradation and help restore the degraded rangelands [13, 19]. After moderate “disturbance”, restoration and rehabilitation, the degraded grassland ecosystems can alter the aboveground community and the soil properties and functions [15, 20]. Dong et al. (2012) [21] showed that the establishment of cultivated grassland on the degraded black soil grasslands in alpine regions of the Qinghai–Tibetan Plateau accelerated the vegetative succession and renewed the soil nutrient cycle, leading to a marked increase in carbon storage. However, it is not clear whether cultivated pasture acts as a CO₂ sink or source.

We sought to identify the influence of the establishment of cultivated grassland on the ecosystem’s carbon budget. We wanted to understand the effects of environmental and biological factors on the carbon budget of the cultivated grassland, but there are few reports on these issues. Therefore, in the present study, we used an eddy covariance system to continuously collect observational data from January 1 to December 31, 2008, and performed a quantitative analysis of the CO₂ flux variations and controlling factors in the TRSR pasture to achieve the following objectives: (1) the quantification of the magnitude of the diurnal and seasonal changes in the net ecosystem CO₂ exchange (NEE) and ecosystem respiration (Rₑ), (2) the
examination of the dependence of carbon fluxes on abiotic and biotic factors and (3) the calculation of the carbon budget of the cultivated pasture during 2008.

Materials and Methods

Study site

The study area is situated in the Geduo pastoral pasture, 25 km southeast of the town of Dawu in Guoluo Prefecture in Qinghai Province, which is located in the TRSR. The geographical coordinates of the area are 100°26′–100°41′E and 34°17′–34°25′N, and the area lies at an elevation of 3980 m above sea level. The area experiences the continental weather typical of the plateau: the annual average sunshine duration is 2576 h, the radiation is strong, there is no absolute frost, and the annual average temperature is −0.5°C. The average temperature in January is −12.7°C, the average temperature in July is 9.8°C, and the annual precipitation (PPT) is approximately 500 mm, with 85% of the PPT concentrated between May and September. The soils are mainly an alpine meadow type and an alpine shrub type. The artificial pasture was established in May 2002, and its total area was 2000 hm². The pasture was sown only with Elymus nutans, and the pasture vegetation height was 40 to 60 cm. During winter, the meadow was subjected to moderate-intensity grazing.

Ethics statement

The study area was owned and/or managed by the Northwest Plateau Institute of Biology of Chinese Academy of Sciences, who gave permission to perform the field research. In the study area, no specific permits were required for collecting samples, and the field studies did not involve endangered or protected species.

Eddy flux and micrometeorological measurements

An eddy covariance flux tower (3.0 m high) was installed at the center of the observation field. The fluxes of CO₂ and H₂O were measured using the eddy covariance method. The uniform fetch was more than 300 m from the tower in all directions. A three-dimensional ultrasonic anemometer, manufactured by Campbell Scientific, Inc. (CSI) (CSAT–3, Logan, UT, USA), was used to measure turbulence. The CO₂ and H₂O densities and the temperature fluctuations were measured using an open-path CO₂/H₂O infrared gas analyzer (CS7500, CSI) and an anemometer–thermometer at 10 Hz, respectively. The average value was output once every 15 min, and the data were saved in a data collection device (CR5000, CSI). The CO₂/H₂O analyzer system was calibrated each year.

Simultaneously with measuring the CO₂ flux, we also measured other routine weather factors. The system for obtaining the routine measurements was installed on the same flux tower as the eddy measurement system. Among these measurements, the net radiation was measured using a net radiometer (CNR–1, Kipp and Zonen, Delft, South Holland, The Netherlands), and the photosynthetic photon flux density (PPFD) was measured using a quantum sensor (LI–190SB, Li–Cor, Lincoln, NE, USA). Both measurements were recorded at a height of 150 cm. The soil temperature was measured using copper–constantan thermocouples (105–T, CSI) at depths of 5, 10 and 30 cm underground. The air temperature (Tₐ) and humidity were measured with a humidity and temperature probe (HUMP45C, CSI) at heights of 110 and 300 cm above the ground. The wind speed and direction were also measured at heights of 110 and 300 cm above the ground using cup anemometers (034A–L and 014A, R. M. Young Co., Traverse, MI, USA). The soil heat flux was measured at a soil depth of 2 cm with heat flux plates (HFT–3, CSI). In total, there were three heat flux plates in the test field, and the average of the soil
heat flux values recorded by the three plates was used. The soil moisture was measured using
time-domain reflectometers (CS615, CSI) at depths of 5, 20 and 30 cm underground. The soil
surface temperature was measured with thermometers (107, CSI) at three points in an area of
1 m². The PPT volume was determined using a tipping bucket (TE525MM, CSI) mounted 70
cm above the ground. The output data consisted of average values calculated every 15 min.
These data were stored in the data collector (CR5000, CSI).

Data processing and energy balance closure

The data were obtained from January 1 to December 31, 2008. All the micrometeorological
and flux data were subjected to data quality control. The raw flux data were preprocessed
before analysis, which primarily included outlier exclusion (±3δ), dimensional coordinate
rotation and the application of the Webb–Pearman–Leuning correction [22], among others.

The surface energy budget of the sample field was examined by performing an ordinary lin-
erar regression (OLR) between the sum of eddy fluxes (LE+H) and the available energy (Rn−G)
during all of 2008: LE+H = 0.69×(Rn−G)+22.06 (R² = 0.84), where LE and H are the latent and
sensible heat fluxes, respectively, Rn is the net radiation, G is the soil heat flux, and all the flux
values are daily averages (MJ m⁻²). Therefore, the energy closure ratio was 69% for the sample
field, and this energy closure slope is within the published energy closure range (0.55 to 0.90)
[23]. Notably, the area contributing to the flux was large, flat and wide open, but the closure
slope was relatively small. This difference might have been caused by the relatively low temper-
atures and wind speeds, and further study is needed to determine whether these weather con-
ditions explain the discrepancy.

Because the flux observation and measurement are affected by the weather conditions at
the site, the data can be processed by eliminating values collected during PPT, morning dew
periods, and nights during the growing season when the carbon flux volume is negative (eco-
system carbon absorption). The data from the night times when the friction wind velocity
(U*)<0.2 m s⁻¹ can be treated as invalid because the turbulence intensity at those times was
not strong enough for the device to properly record CO₂ flux data [24]. After processing, 67%
of the total flux data collected were usable.

The data gaps that were less than 3 h in duration were filled through linear interpolation
between the preceding and subsequent data. The missing data in gaps that exceeded 3 h could
usually be interpolated based on the nonlinear empirical relation between the established car-
bon flux value and the environmental factors [25].

The nighttime data gaps were filled using the soil temperature at a depth of 5 cm (T_soil)
according to Formula (1), and daytime estimates of Rₑ could be obtained according to the
nighttime Rₑ–temperature relationship as follows [26]:

\[ R_e = a \exp(bT_{soil}) \]  

where Rₑ is the nighttime Rₑ rate (μmol CO₂ m⁻² s⁻¹), T_soil is the soil temperature at a depth
of 5 cm, and a and b are fitted coefficients in Formula (1). The temperature sensitivity of respira-
tion (Q₁₀) for the ecosystem was derived from Formula (2), representing the relative growth
volume of the Rₑ for every 10°C temperature increase as follows:

\[ Q_{10} = \exp(10b). \]
The daytime gaps in the CO\(_2\) flux (\(F_c\)) during the growing season were filled through rectangular hyperbolic regression according to Formula (3) [25]:

\[
F_c = \frac{F_{\text{max}} \times \text{PPFD}}{F_{\text{max}} + \alpha \times \text{PPFD}},
\]

where PPFD is the photosynthetic photon flux density (\(\mu\)mol m\(^{-2}\) s\(^{-1}\)), \(F_{\text{max}}\) is the NEE at a saturating light level (\(\mu\)mol CO\(_2\) m\(^{-2}\) s\(^{-1}\)), and \(\alpha\) is the apparent quantum yield (\(\mu\)mol CO\(_2\) \(\mu\)mol\(^{-1}\) photons).

**Vegetation measurement**

The biomass and leaf area index (LAI) were measured six times over the whole growth season. The aboveground biomass measurement adopted the harvesting method, with five randomly collected samples, each including the vegetation within a square area covering 0.25 m\(^2\). The vegetation was cut and brought back to the lab with the roots to be dried in a 65°C thermostatic oven. The LAI was determined from measurements taken with a leaf area meter (LI-3100, Li-Cor). Based on plant phenology data from the TRSR, we assumed that the biomass and the LAI for the sampling field were both zero before April 20 (day of year (DOY) 111) and after October 18 (DOY 292), which marked the beginning and end of the growing season, respectively. LAI gaps were linearly interpolated to daily intervals [27].

**Path analysis**

The relationships between the CO\(_2\) flux and environmental factors were evaluated using path analysis. Path analysis is the continuation of the simple correlation coefficient and decomposes the correlation coefficient based on multiple regression. It uses direct and indirect paths to indicate the direct effect of a variable on a dependent variable and the indirect effects of other variables on a dependent variable [28, 29]. Path analysis was performed using the standardization of multiple linear regression models that are included with the RGE and CORR packages of SAS 9.4 software (SAS Institute Inc., Cary, NC, USA).

The decision coefficient \(R^2(j)\) is often used to quantify the integrated determination effect of environmental factors (\(x_j\)) on the ecosystem CO\(_2\) flux (\(y\))[30]:

\[
\begin{align*}
R^2_{(j)} &= R^2_J + \sum_{j\neq i} R^2_{ji} \\
R^2_J &= b^2 \\
R^2_{ji} &= 2b_jr_{ji}b_i
\end{align*}
\]

\(R^2(j)\) contains not only the direct determination effect \(R^2_J\) of \(x_j\) only but also the indirect determination coefficient (\(\sum R^2_{ji}\)) related to \(x_j\). Additionally, \(b^2\) represents the path coefficient, and \(r_{ji}\) represents the correlation. The \(x\) corresponding to the maximum value of \(R^2_{(j)}\) has the maximum synthesis effect on \(y\) and is called the main decision-making factor. In contrast, the \(x\) corresponding to the minimum value of \(R^2_{(j)}\) is called the main confined factor.

**Results**

**Meteorological and biological factors**

Fig 1a shows that the PPFD peaks occurred between May and August, when the solar elevation angle was higher than in other seasons. Because of the high elevation of the plateau, the PPFD
Fig 1. Temporal variation during 2008 in the (a) photosynthetic photon flux density (PPFD) and vapor pressure deficit (VPD), (b) daily mean air temperature ($T_a$) and soil temperature at a depth of 5 cm ($T_{soil}$), (c) daily precipitation (PPT) and soil water content (SWC) at a depth of 5 cm, (d) daily net ecosystem CO$_2$ exchange (NEE) and ecosystem respiration ($R_e$) and (e) leaf area index (LAI).

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values of the plateau also tended to be high, and the maximum daily values reached 695.9 μmol m⁻² s⁻¹ (Fig 1a). The vapor pressure deficit (VPD) also showed significant seasonal variation, reaching its highest and lowest values, approximately 1.32 and 0.04 kPa, respectively, during the growing season and during the winter (Fig 1a). The daily mean Tₐ and Tₛ values showed the same seasonal variation trends, ranging from −17.8 to 12.1°C for Tₐ and from −8.4 to 16.5°C for Tₛ. The annual average was −0.54°C for Tₐ and 4.2°C for Tₛ (Fig 1b). The annual PPT was 628.9 mm, which was higher than the average PPT across multiple years (approximately 500 mm), and the PPT during May–September accounted for 66.4% of the annual PPT. After October, the PPT was significantly reduced (Fig 1c). The variation in the soil water content (SWC) was strongly dependent on the PPT; the SWC was higher from May to October than at other times of the year and was generally maintained above 20% (Fig 1c).

Table 1 shows that the LAI for the field in the sampled pasture started to increase at the end of April and reached the maximum LAI (2.9±0.3) at the end of August. In September, the LAI decreased rapidly because of leaf aging. The growing season for the pasture in 2008 (DOY 113–292) could be divided into the following four periods (Table 1): the beginning growing period (I, DOY 113–145), the fast growing period (II, DOY 146–194), the peak growing period (III, DOY 195–252) and the aged growing period (IV, DOY 253–292).

Diurnal course of CO₂ exchange

Fig 2d shows that the daily variation in the NEE was regular during every growth period. This variation was most likely a reflection of daytime absorption and nighttime emission. In the
Fig 2. Average diurnal cycles of the (a) photosynthetic photon flux density (PPFD), (b) air temperature ($T_a$), (c) vapor pressure deficit (VPD), (d) net ecosystem CO$_2$ exchange (NEE) and (e) ecosystem respiration ($R_e$) in different periods of the growing season. Bars indicate ± standard error (SE). The time zone is Beijing Standard Time (BST).

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morning, the NEE was converted from a positive value (representing carbon emission) to a negative value (representing carbon absorption). The absorption value reached its maximum before noon (10:00–11:00 h) and then started to diminish. Near evening (approximately 19:00 h), the NEE changed from a negative value to a positive value. The hourly maximum and minimum NEE rates, which were 3.25 and −7.89 μmol CO₂ m⁻² s⁻¹, respectively, both occurred during the peak growing period. Fig 2e shows that the hourly maximum Rₑ rate of the pasture, which occurred at approximately 16:00 h during the peak growing period, was 5.03 μmol CO₂ m⁻² s⁻¹.

**Seasonal course of CO₂ exchange**

Fig 1d shows that from January until the end of April, the pasture NEE was greater than 0 because the aboveground vegetation had withered; thus, the ecosystem was emitting carbon (NEE>0). Starting on May 1 (DOY 121), as the vegetation began to appear, the NEE began to drop to below 0. The whole ecosystem converted from carbon emission to carbon absorption (NEE<0) and reached peak carbon absorption between July and August. Starting in September, as the vegetation aged, the carbon absorption capability of the pasture gradually degraded. By the end of October, the NEE began to exceed 0, and the whole ecosystem engaged in carbon emission (NEE>0). The maximum daily absorption value, −2.91 g C m⁻² day⁻¹, occurred on August 12 (DOY 225). The ecosystem appeared to be a carbon sink during May–October (NEE<0), with the maximum CO₂ uptake occurring in August, i.e., −50.59 g C m⁻² month⁻¹ (Fig 3). The annual NEE for the pasture in 2008 was −140.04 g C m⁻² year⁻¹. Thus, the pasture was a carbon sink during 2008.

There is a significant seasonal change in Rₑ for the cultivated pasture. Fig 3 shows an increasing trend in CO₂ emissions from winter (January) to summer (July–August) and a decreasing trend until autumn (September), with maximum and minimum values in July (90.36 g C m⁻² month⁻¹) and January (2.96 g C m⁻² month⁻¹), respectively (Fig 3). In 2008, the daily maximum Rₑ for the pasture was 5.04 g C m⁻² day⁻¹ on July 28 (DOY 210) (Fig 1e)). The annual Rₑ was 403.57 g C m⁻² year⁻¹ in 2008, of which approximately 85.5% fell in the growing season, from May to September.

**The relationship between the daytime NEE and the PPFD**

We used Formula (3) to depict the relationship between the daytime NEE and the PPFD. The NEE data were averaged using PPFD bins of 100 μmol m⁻² s⁻¹. As shown in Fig 4, at PPFD<1600 μmol m⁻² s⁻¹, the daytime NEE decreased as the PPFD increased. However, for PPFD>1600 μmol m⁻² s⁻¹, the daytime NEE increased as the PPFD increased (Fig 4). Therefore, Formula (3) was only valid for depicting the relationship between the NEE and the PPFD for PPFD<1600 μmol m⁻² s⁻¹. During the entire growing season, the model-derived α and F_max values in the pasture increased as the canopy developed, and their maximum values occurred during the peak growing period, reaching −0.0358 μmol CO₂ μmol⁻¹ photons and −8.69 μmol CO₂ m⁻² s⁻¹, respectively (Table 1). During the entire growing season, the α and F_max values in the pasture were −0.0275 μmol CO₂ μmol⁻¹ photons and −7.86 μmol CO₂ m⁻² s⁻¹, respectively.

To further study the influence of environmental factors on the NEE–PPFD curve, we inspected the NEE–PPFD curves generated under different Tₛ conditions (Tₛ<5˚C, 5˚C<Tₛ≤15˚C, and Tₛ>15˚C), SWC conditions (SWC≤25%, 25%<SWC≤30%, and SWC>30%) and VPD conditions (VPD≤0.6 kPa and VPD>0.6 kPa). Under the aforementioned micrometeorological conditions, the NEE could be further subdivided based on the PPFD (using 100 μmol m⁻² s⁻¹ PPFD subdivisions), and the NEE was then averaged for each
PPFD level. Statistically, this method can reduce or offset the errors that occurred during measurement [25].

In the pasture, the $F_{\text{max}}$ and $\alpha$ values for the NEE–PPFD curve were under the influence of the SWC, $T_a$ and the VPD. Both $F_{\text{max}}$ and $\alpha$ increased as the SWC increased, and at a SWC $< 20\%$, $F_{\text{max}}$ and $\alpha$ were significantly lower than they were when the SWC $> 30\%$. $F_{\text{max}}$ and $\alpha$ were highest when $5^\circ C < T_a < 15^\circ C$. $F_{\text{max}}$ and $\alpha$ decreased as the VPD increased, and at VPD $> 0.6$ kPa, $F_{\text{max}}$ and $\alpha$ were 93% and 38%, respectively, of their values at VPD $\leq 0.6$ kPa (Fig 5 and Table 1).

The relationships between the daytime NEE and the values for $T_a$, VPD and SWC

For statistical purposes, the daytime NEE data were averaged based on abiotic controls divided into bins, with bin widths of 1°C for $T_a$, 1% for SWC and 0.1 kPa for VPD over all the PPFD values. Fig 6a and 6b show that the relationships between the daytime NEE and $T_a$ and the VPD can be depicted by a quadratic function, and the stepwise regression analysis results

Fig 3. Seasonal variation in the monthly net ecosystem CO$_2$ exchange (NEE) and in the ecosystem respiration ($R_e$) in the studied cultivated pasture ecosystem in 2008.

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indicated that the optimal $T_a$ and VPD values for the pasture ecosystem were 14.1˚C and 0.65 kPa. The daytime NEE decreased as the SWC increased (Fig 6c).

For the entire growing season, the nighttime NEE data were bin averaged using $T_{soil}$ bins of 1˚C. Fig 6 shows that $R_e$ increased exponentially as the temperature increased, and the $Q_{10}$ for the pasture during the entire growth season was 3.0 (Fig 7 and Table 2), with values of 1.9, 2.9,
Fig 5. Relationship between the net ecosystem CO₂ exchange (NEE) and PPFD for different values for (a) air temperature ($T_a$), (b) soil water content (SWC) and (c) vapor pressure deficit (VPD). The daytime NEE data were averaged over PPFD bins of 100 μmol m⁻² s⁻¹. Bars indicate ±SE. Formula (3) was used to fit the data when the PPFD was below 1,600 μmol m⁻² s⁻¹; the regression coefficients are presented in Table 1.

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Fig 6. Relationships between the daytime NEE and (a) air temperature ($T_a$), (b) vapor pressure deficit (VPD) and (c) soil water content (SWC) at a depth of 5 cm. The daytime NEE data were averaged over bin widths of 1°C for $T_a$, 0.1 kPa for VPD and 1% for SWC. Bars indicate ±SE. The dotted lines in (a) and (b) were fitted using a piecewise regression model.

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1.8 and 2.7 for the beginning, fast, peak and aged growing periods, respectively (Fig 7 and Table 2).

To further investigate the effect of the SWC on Tsoil and Re, we investigated the Tsoil–Re relation under different SWC conditions (SWC ≤ 25%, 25% < SWC ≤ 30% and SWC > 30%). The results showed that the Q10 of the cultivated grassland reached its maximum at 25% < SWC ≤ 30%.

The relationship between the daily NEE and the LAI

Fig 8 indicates that during the entire growing season, the cultivated grassland’s daily integrated NEE and LAI showed a linear relationship in which NEE = (−0.449 ± 0.005) × LAI − (0.291 ± 0.008), n = 175, adjusted R² = 0.325 and F = 84.8. Therefore, 32.5% of the variation in the NEE could be explained by variation in the LAI.

Path analysis evaluation of the daily NEE

The path analysis results show that during the entire growing season, the relationship between the cultivated grassland’s daily integrated NEE and environmental factors can be described using the following formula: NEE = 2.22778 + 0.46160 T + 0.34505 × LAI − 0.18 T × SWC + 2.11380 × VPD, with n = 1071, adjusted R² = 0.6025 and F < 0.0001. The direct path coefficients (R²j) of the environmental factors (xj) affecting the daily integrated NEE (y) are ranked as follows: Tsoil (0.46) > VPD (0.21) > SWC (−0.20) > LAI (−0.31) > Tα (−0.53) > PPFD (−0.59) and for the decision coefficient R²(j), Tsoil (0.12) > SWC (−0.18) > LAI (−0.20) > VPD (−0.34) > Tα (−0.44) > PPFD (−0.69) (Fig 9).

Discussion

The effect of biotic and abiotic controls on the NEE

The maximum Fmax for the pasture ecosystem (−8.69 μmol CO₂ m⁻² s⁻¹) occurred during the peak growing period (Table 1) and was nearly identical to the Fmax for a steppe–Kobresia meadow during the peak growing period (−8.7 μmol CO₂ m⁻² s⁻¹) [11]. However, the pasture maximum Fmax was lower than the values reported for other grassland ecosystems (from −9.6 to −40.2 μmol CO₂ m⁻² s⁻¹) [27]. For the entire growing season, the α value for the pasture was 0.02754 μmol CO₂ μmol⁻¹ photons, which was higher than that for the steppe–Kobresia meadow (−0.0159 μmol CO₂ μmol⁻¹ photons) [11]. However, the α for the pasture for the entire growing season was at a moderate to low level compared with the values for other grassland and cropland ecosystems, as reported by Li et al. (2005) [27] (from −0.008 to −0.465 μmol CO₂ μmol⁻¹ photons). These findings indicate that the light-use efficiency of the pasture was low. This low efficiency was related to the use of C3 vegetation to establish the pasture and to the high elevation and low temperatures of the pasture ecosystem [31].

Under low Tα values (Tα ≤ 5˚C), the Fmax and α of the NEE–PPFD curve for the pasture were relatively low (Table 1), mainly because low temperature can suppress the activity of photosynthesis-related enzymes [32]. This situation was also observed in the desert steppe of Inner Mongolia [33]. At SWC < 20%, the Fmax and α values of the NEE–PPFD curve for the pasture were relatively low (Table 1), primarily because the low SWC can constrain plant growth. A similar situation also occurred for the steppe [27]. During the aged growing period, the Fmax and α values of the NEE–PPFD curve for the pasture decreased significantly. This finding is related to the reduced chlorophyll content of older plants, which also show decreased activity of photosynthesis-related enzymes [34]. Similar results have also been reported for a fenced steppe [30].
Fig 7. Relationship between the nighttime net ecosystem CO$_2$ exchange (NEE) and soil temperature ($T_{soil}$) for (a) the entire growing season, (b) the different growing periods and (c) different soil water content (SWC) ranges. The nighttime NEE data were averaged over a bin width of 1°C for $T_{soil}$. Bars indicate ±SE. Formula (1) was used to fit the data; the regression coefficients are presented in Table 2.

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The optimal $T_a$ for CO$_2$ uptake in the pasture was $14.1{^\circ}C$ (Fig 5a), which was quite similar to that of an alpine meadow ($15{^\circ}C$) [35]. In our study, $T_a$ had marked effects on the NEE. The NEE decrease at lower temperatures was most likely caused by the slow growth rate during the early and late stages of the growing season, whereas the depression of the NEE at relatively higher temperatures could be ascribed primarily to enhanced respiration and depressed plant photosynthesis in response to high temperatures and high radiation levels [36].

In many ecosystems, moisture is an important factor that influences the daytime NEE. The daytime NEE for the pasture decreased with increases in the soil moisture (Fig 5c). This trend indicated that increased soil moisture can improve the carbon absorption capability of the pasture. Similar results have been reported for a Mongolian steppe [27]. Preliminary studies demonstrated that a lack of moisture could result in the closure of plant stomata, further reducing plant CO$_2$ absorption. In addition, stomatal closure had a significant effect on leaves. An increasing leaf temperature can enhance leaf photorespiration, which further reduces CO$_2$ acquisition by the plants.

In the present study, the daytime NEE and the VPD of the pasture were quadratically related (Fig 5b), and similar results were observed for a temperate desert steppe [33]. The daytime suppression of the NEE by a high VPD could be primarily attributed to the physical

![Graph](image-url)

**Fig 8.** Relationship between the daily net ecosystem CO$_2$ exchange (NEE) and the leaf area index (LAI) from April 20 to October 18, 2008.

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**Table 2. Parameters describing the characteristics of the relationship between the nighttime NEE and $T_{soil}$ (Formulas (1) and (2)).**

| Treatment                      | DOY   | SWC     | a     | b     | $R^2$ | $Q_{10}$ | P value |
|--------------------------------|-------|---------|-------|-------|-------|----------|---------|
| Beginning growing period      | 113–145 | 23.17±2.13 | 0.4661 | 0.0664 | 0.66  | 1.9425   | <0.0001 |
| Fast growing period           | 146–194 | 26.08±1.74 | 0.4809 | 0.1068 | 0.94  | 2.9096   | <0.0001 |
| Peak growing period           | 195–252 | 35.62±2.23 | 1.2801 | 0.0583 | 0.83  | 1.7914   | <0.0001 |
| Aged growing period           | 253–292 | 37.63±3.02 | 0.6434 | 0.0992 | 0.89  | 2.6966   | <0.0001 |
| Entire growing season         | 113–292 | 32.41±2.25 | 0.5551 | 0.1088 | 0.95  | 2.9683   | <0.0001 |
| SWC≤25%                       |       | 0.3136   | 0.1304 | 0.96  | 3.6840 | <0.0001 |
| SWC>30%                       |       | 0.3188   | 0.1438 | 0.79  | 4.2123 | <0.0001 |
| SWC>30%                       |       | 0.6936   | 0.0960 | 0.91  | 2.6117 | <0.0001 |

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relationship between the temperature and the VPD (Fig 2). Because this relationship can affect the hydraulic status of plants and leaves, leading to leaf closure, it can affect the acquisition of CO₂ by plants [37].

The process of carbon exchange between plants and the atmosphere is jointly regulated by multiple environmental factors (such as the PPFD, Tₐ, SWC and VPD); thus, it is difficult to identify a specific effect on the NEE caused by a single factor, especially between Tₐ and the VPD, given that a rising Tₐ is always associated with an increased VPD. Therefore, future studies of the response mechanisms of the NEE to environmental factors should use both modeling and multivariate analysis.

The effect of Tₛ and the SWC on the nighttime NEE

Rₑ is affected by multiple environmental and biological factors, and Tₛ and the SWC can be regarded as controls [38].

Rₑ shows an exponential function with increasing temperature [39]. During the entire growing season in the present study, the Q₁₀ for the pasture was 3.0 (Fig 7a and Table 2), which was higher than the Q₁₀ value for low-elevation grassland ecosystems around the world (2.1 according to Zheng et al., 2009 [40]). A previous study showed that the Q₁₀ value for Rₑ decreases as the temperature increases [40]. Here, the relatively high Q₁₀ value for the TRSR could result from the low temperature on the plateau. Therefore, the results of the current study indicated that in the context of global warming, the TRSR pasture has a relatively strong carbon emission potential.
The $Q_{10}$ value reached its maximum at a medium SWC (Fig 7c and Table 2). This situation also occurred in a Stipa krylovii steppe [41]. At a high SWC, the soil moisture can hinder the diffusion of $O_2$. Therefore, a high SWC can suppress the decomposition of organic matter and decrease the microbial respiration rate. Under these conditions, the CO$_2$ release and temperature are not sensitive, and the $Q_{10}$ value is relatively low. However, at a low SWC, the primary component, composed of R$_e$, derives from the more recalcitrant carbon material, and the $Q_{10}$ of this material is low [42]. The situations discussed above cannot explain the high $Q_{10}$ value during the aged growing period (SWC > 30%). This result indicated that plant phenology was another factor affecting the $Q_{10}$. A similar result has also been reported for a desert steppe [33].

The effect of the LAI on the daily NEE

The structure of the plant canopy, especially the leaf area and light interception capability, determine the quantity of radiation absorbed and reflected by the plant canopy. Therefore, these factors can have a direct influence on plant photosynthesis [43, 44]. For the pasture ecosystem, the LAI could explain 32.5% of the NEE variation (Fig 8), and this percentage of explained variation was higher than that found for a desert steppe (26%, [33]). This finding could be attributed to the additional PPT received by the pasture ecosystem during the growing season.

Effects of environmental factors on the CO$_2$ flux

In the present study, path analysis was conducted to evaluate the relationships between various environmental factors and the NEE in the studied pasture during the growing season (Fig 9). Among the six factors directly affecting the NEE, the path coefficient for $T_{soil}$ in the pasture was 0.46, which was much higher than the contributions of the other factors. $T_{soil}$ explained most of the variability in the daily average NEE in the pasture. This conclusion agrees with the conclusions of Wang et al. (2011) [30]. Temperature is an important factor regulating several ecological processes and properties associated with the CO$_2$ flux and its response to climate variation, including the evapotranspiration rate [45], canopy development [46] and SWC [47]. However, the minimum value of the decision coefficient $R^2$ (the synthesis index), which indicates the importance of the PPFD on the NEE, was −0.69 in the pasture. Therefore, the PPFD was the main factor, although it must be limited to lower values for dynamic ecosystem CO$_2$ uptake. Similar results have been found for subalpine environments [28] because light is the most important ecological factor for regulating plant photosynthesis; both high and low light intensity can limit plant absorption and fix CO$_2$ [48].

On the one hand, due to the high altitude in the TRSR, the PPFD was often greater than 1600 $\mu$mol m$^{-2}$s$^{-1}$ at around noon on clear days, exhibiting a well-defined inhibitory effect on plant photosynthesis (Fig 4). On the other hand, the growing season on the Qinghai–Tibet Plateau coincides with precipitation and heat, and 69% of the days during the growing season of 2008 were rainfall days (Fig 1c). Constant rainfall reduces the time and intensity of sunlight. According to the statistics, the ecosystem daily cumulative PPFD value was less than 30 mol m$^{-2}$ day$^{-1}$, accounting for 31% of the total growing season days (Fig 1c). According to Satio et al., (2009) [29], low PPFD values (daily PPFD cumulative values less than 30 mol m$^{-2}$ day$^{-1}$) are important photosynthesis inhibition factors for Qinghai–Tibetan grassland ecosystems. Moreover, Satio et al. [29] also found that the most conducive daily PPFD cumulative value for plant photosynthesis is approximately 50 mol m$^{-2}$ day$^{-1}$ on the Qinghai–Tibetan Plateau (Fig 1b), although this phenomenon only occurs over 5% of the growing season (Fig 1b). Therefore, the special geographical environment
(high altitude) and climatic conditions (precipitation and heat simultaneously) are the main reasons for the PPFD, becoming the main factor constricting NEE in the ecosystem.

Diurnal and seasonal variation in the NEE and $R_e$

At various stages during the growing season, the carbon absorption of the pasture ecosystem was significantly stronger at noon than before noon, indicating that the daily NEE is significantly suppressed around noon (Fig 2d). Fu et al. (2006) [36] also obtained a similar asymmetrical distribution of the NEE in a study of an alpine shrub. Because of photosynthetic depression at high temperatures, as well as stomatal closure at high PPFD levels, the carbon assimilation was severely restricted at noon and during the early afternoon. For most plants on the Qinghai–Tibetan Plateau, the photosynthetic depression at noon is a common phenomenon. This response of the plants is primarily due to enhanced respiration and depressed photosynthesis at high temperatures under high-radiation conditions [36].

For the pasture ecosystem, the magnitude of the maximum hourly NEE was $-7.89 \mu \text{mol CO}_2 \text{m}^{-2} \text{s}^{-1}$, and this value was lower than that of other grassland ecosystems located at similar latitudes, such as the tall prairie grassland native to North America ($-23 \mu \text{mol CO}_2 \text{m}^{-2} \text{s}^{-1}$) [49], other prairie plains in the USA ($-19.5 \mu \text{mol CO}_2 \text{m}^{-2} \text{s}^{-1}$) [50], and the alpine meadow at Haibei Station ($-10.8 \mu \text{mol CO}_2 \text{m}^{-2} \text{s}^{-1}$) [51]. This lower magnitude of the NEE in the current study could be attributed to the relatively low temperature of the pasture ecosystem and to the C3 composition of most of the pasture vegetation.

For the pasture ecosystem, the magnitude of the maximum daily NEE was $-2.91 \text{g C m}^{-2} \text{day}^{-1}$, which is at the lower end of the maximum daily NEE variation range (from $-1.91$ to $-9.3 \text{g C m}^{-2} \text{day}^{-1}$) for other grassland ecosystems [27]. This relatively low value was related to the practice of single-species sowing, which greatly reduces the plant diversity [52]. Naeem et al. (1994) [53] found that a reduction in plant diversity can cause a simplification in the canopy structures and a reduction in the light acquisition and utilization efficiency of a plant colony, thereby reducing the CO$_2$ uptake of the overall ecosystem.

The annual NEE of the pasture in 2008 indicated that the pasture acted as a medium-strength carbon sink compared with other grassland ecosystems (from $-18$ to $-274 \text{g C m}^{-2} \text{year}^{-1}$) [27]. The low $T_a$ in the environment and the matching traits of the cool-adapted plants (e.g., the depression of the NEE at relatively higher temperatures) might have operated as important environmental restrictions on the potential of the pasture to act as a carbon sink.

The plant pasture showed a significant seasonal change in CO$_2$ emissions, with the maximum value occurring during the peak growing period (Fig 3). Similar patterns have been found for a native alpine meadow [16]. Li et al. (2015) [16] attributed this change to high temperatures and biomass during the peak growing period. For the pasture ecosystem, the annual $R_e$ in 2008 was 403.57 g C m$^{-2}$ year$^{-1}$, which was lower than that in a native alpine meadow near Haibei Station (488.5–555.6 g C m$^{-2}$ year$^{-1}$) [16, 54] and was similar to or less than other values (from 138 to 2392 g C m$^{-2}$ year$^{-1}$) for other grassland ecosystems [55]. This finding indicates that the construction of cultivated grassland will not increase CO$_2$ emissions in the TRSR.

Conclusion

We adopted eddy covariance to investigate the NEE for a single-sown cultivated pasture of *Elymus nutans* in the TRSR in 2008. Our results show that for the pasture, the NEE for the entire year was 140.01 g C m$^{-2}$ year$^{-1}$. Therefore, the cultivated grassland was a carbon sink during 2008. Because of the low temperatures in the TRSR, the annual $R_e$ of the pasture was only 403.57 g C m$^{-2}$ year$^{-1}$ in 2008, lower than those in most grassland ecosystems around the
world. This finding implies that cultivated grassland establishment can both effectively resolve the grass–livestock conflict and properly improve the grassland carbon fixation capacity in the TRSR. Moreover, the Q₁₀ value in the pasture for the entire growing season was 3.0, which was higher than that in low-elevation grassland ecosystems around the world, indicating greater sensitivity to elevated temperatures in the future in terms of ecosystem carbon loss in the study area. During the daytime, the NEE was primarily regulated by the PPFD; at night, the NEE was mainly regulated by T_soil. A higher temperature can suppress photosynthesis in pastures, reducing the carbon absorption capacity of pasture ecosystems. The daily NEE and LAI were linearly related, and 32.5% of the NEE variation can be interpreted based on the LAI variation. Path analysis showed that the daily NEE in the growing season of cultivated grassland was controlled by various ecological factors at the same time. Among them, T_soil was found to be the main decision-making factor for the daily NEE, and PPFD is the main constraining factor of the NEE in the studied ecological system.

Supporting Information

S1 File. Language polish editing certificate.
(PDF)

S1 Table. Temporal variation in meteorological and biological factors and CO₂ exchange during 2008. Abbreviations: PPFD, photosynthetic photon flux density; VPD, vapor pressure deficit; T_a, daily mean air temperature; T_soil, soil temperature at a depth of 5 cm; PPT, daily precipitation; SWC, soil water content at a depth of 5 cm; NEE, daily net ecosystem CO₂ exchange; Rₑ, ecosystem respiration; and LAI, leaf area index (LAI).
(XLSX)

S2 Table. Diurnal variation in meteorological and CO₂ exchange (PPFD, T_a, VPD, Rₑ and NEE). Abbreviations and symbols are as described for S1 Table.
(XLSX)

S3 Table. The relationship between the daytime NEE and the PPFD. Abbreviations and symbols are as described for S1 Table.
(XLSX)

S4 Table. Relationship between the NEE and environmental factors. Abbreviations and symbols are as described for S1 Table.
(XLSX)

S5 Table. Relationship between Rₑ and T_soil. Abbreviations and symbols are as described for S1 Table.
(XLSX)

S6 Table. Path coefficients between the NEE and environmental factors in the studied pasture. Abbreviations and symbols are as described for S1 Table.
(XLSX)

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