Comparative Sensitivities of Zooplankton to Ocean Acidification Conditions in Experimental and Natural Settings

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Zooplankton can serve as indicators of ecosystem health, water quality, food web structure, and environmental change, including those associated with climate change and ocean acidification (OA). Laboratory studies demonstrate that low pH and high pCO₂ associated with OA can significantly affect the survival, growth, calcification, development, and reproduction of planktonic organisms (Kroeker et al., 2013; Busch et al., 2014; Cripps et al., 2014; Busch and McElhany, 2016; McLaskey et al., 2016; Tills et al., 2016; Bednarsk et al., 2017; Lischka and Riebesell, 2017), and that sensitivity to current and projected levels of pH

INTRODUCTION

Zooplankton are excellent indicators of ecosystem health, food web function, and water quality (Beaugrand, 2005) because their short life cycles allow them to respond to seasonal variations and abrupt changes in environmental conditions. While the response of zooplankton to ocean acidification (OA) is understudied compared with that of many benthic species, the critical importance of zooplankton to marine food webs underscores the need to better understand their sensitivities to changing seawater carbonate chemistry. Laboratory studies show that low pH and high pCO₂ associated with OA can significantly affect the survival, growth, calcification, development, and reproduction of planktonic organisms (Kroeker et al., 2013; Busch et al., 2014; Cripps et al., 2014; Busch and McElhany, 2016; McLaskey et al., 2016; Tills et al., 2016; Bednarsk et al., 2017; Lischka and Riebesell, 2017), and that sensitivity to current and projected levels of pH
and pCO$_2$ can vary substantially among zooplankton taxa. Understanding the ecological context in which these responses occur is important, given the co-occurring environmental stressors (e.g., increasing temperature and oxygen stress) to which zooplankton are exposed.

Here, we present the results of an exploratory study to assess the level of correspondence between the results of experimental studies and field observations of zooplankton abundance in a natural setting. Specifically, we asked whether relative sensitivities to pCO$_2$ as reported in the literature and revealed through meta-analysis could explain a significant proportion of the variation in zooplankton abundance at two time points, across six stations, in Puget Sound, Washington. We focused on Puget Sound because pH and pCO$_2$ are known to vary substantially across short spatial and temporal scales, often reaching values that are more extreme than in open-ocean environments (Feely et al., 2010; Bianucci et al., 2018; Pelletier et al., 2018), and because the fate of zooplankton in nearshore environments such as this imposes critical controls on food webs and fisheries in coastal environments. We interpret our results in the context of long-term monitoring studies and their utility in detecting biological response to ocean change.

EVIDENCE FROM CONTROLLED EXPERIMENTS

To better understand zooplankton sensitivities to OA conditions in Puget Sound, we performed a meta-analysis of experimental studies reporting responses of seasonally abundant, holoplanktonic taxa known to occur in Washington State’s marine waters. Meroplanktonic larvae of shellfish, including crabs, were also included due to their regional economic and ecological importance. Studies published in peer-reviewed journals over the 10-year period from January 2009 to February 2019 were included. We modified the approach of Wittmann and Pörtner (2013) to determine the effect of increasing pCO$_2$ on major taxonomic groups and to compare sensitivity across groups. The coarse nature of our taxonomic groupings was necessitated by the relative scarcity of published results and was similar to the broad taxonomic levels used by Wittmann and Pörtner (2013). For each species, we interpolated and extrapolated data to help compensate for missing values and cleaned the dataset to eliminate duplicative response categories. For example, because weight and diameter are both measures associated with growth, only one was used in the meta-analysis, thereby reducing the likelihood of over-representation of response categories by including only one experiment per response category per study. Thirty-six studies met our criteria for inclusion in the meta-analysis, representing 17 species and eight taxonomic groups.

Negative responses increased with increasing levels of pCO$_2$ across taxonomic groups, with the exception of shrimp and larvaceans (Figure 1). Pteropods emerged as the taxon most sensitive to OA, showing negative effects at pCO$_2$ levels as low as 530 µatm. Copepods were the second-most sensitive group, with reproduction and survival negatively affected at pCO$_2$ levels as low as 824 µatm. Larval bivalve calcification, growth, development, reproduction, and survival were negatively affected, with calcification affected at pCO$_2$ levels as low as 545 µatm. Shrimp were the only taxon to show positive response to OA conditions, but this was not consistent across studies. Krill development, growth, metabolism, and survival were all negatively affected at pCO$_2$ levels as low as 956 µatm, while crab larvae showed tolerance up to pCO$_2$ levels of 1,361 µatm. Jellyfish growth was affected at pCO$_2$ levels above 1,000 µatm, but other response categories showed no effect. Larvaceans showed the least sensitivity, with no significant response to the levels of pCO$_2$ tested. We note, however, that our meta-analysis included only one study of larvaceans.

Pooling data across all taxa revealed that calcification was the response category most sensitive to increasing pCO$_2$, followed by development, growth, reproduction, survival, and metabolism, in that order.

MATCHING EXPERIMENTAL DATA WITH FIELD OBSERVATIONS

To compare results of the meta-analysis with field observations, we collected zooplankton via vertical tows and used standard techniques to characterize water chemistry at six stations in Puget Sound in 2017 from June 23 to 30 and again from August 25 to September 01 (hereafter referred to as June and August). The sampling stations and dates were chosen to cover a range of physical and chemical conditions, targeting stations where high pCO$_2$ or low pH had been previously reported (Feely et al., 2010; McLaskey et al., 2016; Pelletier et al., 2018). We chose abundance as our response variable because abundance or its co-variate, biomass, is the most commonly measured variable in long-term monitoring studies, including those that have been used to detect changes due to ocean condition (e.g., Hays et al., 2005; Mackas and Beaugrand, 2010; McKinstry and Campbell, 2018). Moreover, abundance has been used to detect response to change in short-term studies of marine microbes (Currie et al., 2017; Rahlf et al., 2021).

Twenty-five taxonomic groups of zooplankton were represented in our samples. Copepods were abundant across stations, accounting for 70% of total abundance across all samples, while larvaceans were abundant at two stations in August, accounting for 15% across all samples.

To compare sensitivities between experimental and field settings, we used the taxon-specific sensitivities revealed by the meta-analysis to infer a hierarchical range of susceptibilities to pCO$_2$. We grouped taxa into four categories, with pteropods (and all other gastropods), copepods, and bivalve larvae designated as “most susceptible”; shrimp and krill as “susceptible”; crab larvae, jellyfish, and larvaceans as “least susceptible”; and all other taxa as “insufficient data.” The distribution of differentially susceptible taxa varied across stations and sampling points (Figure 2), but overall we detected no clear association between zooplankton abundance and sensitivity as inferred from meta-analysis.

The BIOENV function (R package “vegan”: Clarke and Ainsworth, 1993) was used to estimate the combination
of environmental variables most highly correlated with the biological matrix. The best fit model did not include pCO$_2$ or its covariate, pH, but instead included a combination of minimum temperature, maximum dissolved oxygen, and minimum fluorescence. Research indicates that zooplankton are negatively affected by pCO$_2$ conditions that currently occur in Puget Sound, including conditions that were observed in this study (Barton et al., 2012; McLaskey et al., 2016;
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**FIGURE 2** Total abundance of zooplankton (individuals/m$^3$) by broad taxonomic grouping at six stations established by the Puget Sound Regional Synthesis Model program (P8, P12, P15, P28, P38, and P402) sampled in Puget Sound, WA, United States in June and August 2017. Isopods and arachnids were recorded three or fewer times and were omitted. The color of each taxon is based on estimated levels of susceptibility to ocean acidification derived from the meta-analysis: red taxa are most susceptible, yellow taxa are susceptible, blue taxa are least susceptible, and gray taxa have insufficient data. The mean pH for each station is recorded in white in the center of each pie.

Bednaršek et al., 2017, 2020c). Our results, however, did not suggest that zooplankton abundance was responsive to low pH or waters undersaturated with respect to $Q_c$; if effects were present, they were not detectable by our methodology. The mean values of pH observed across stations ranged from 7.45 to 7.90; for comparison, the pH predicted for the surface ocean by the end of this century is 7.75 (IPCC, 2014).

**DISCUSSION**

Zooplankton abundance as measured in the field did not reflect evidence from laboratory studies regarding organismal response to low pH/high pCO$_2$. Taxa reported to be most susceptible to low pH in the laboratory reached high abundances at stations with pH as low as 7.45. Our analysis revealed that temperature, dissolved oxygen, and fluorescence, but not pH, were leading environmental variables influencing zooplankton abundance at the locations and dates sampled.

Zooplankton abundance tends to be positively associated with primary production (or its correlate, chlorophyll fluorescence), and our observations confirm the importance of primary production to zooplankton abundance in Puget Sound. This strong positive relationship is consistent with evidence that food availability can mediate organismal response to OA (Thomsen et al., 2013; Pansch et al., 2014; Ramajo et al., 2016). We note, however, that this is not always the case: for example, Brown et al. (2018, 2020) found that food sufficiency exacerbated the negative effects of low pH on organismal performance. Given our limited dataset, the relationship between food availability and zooplankton abundance in our study should be interpreted with caution with respect to zooplankton response to OA.

Although several of our stations were similar in the physical and chemical properties we measured, their zooplankton communities differed, especially in our samples from August. Notably, the stations differ in depth, but according to our analyses, depth did not appear as a significant explanatory factor. Factors that we did not measure, such as life-cycle
such study demonstrates that early life stages of a common underestimate susceptibility among the taxa we included. One of our meta-analysis suggests that our results could over- or response or for use as input to numerical models. against the use of coarse taxonomic data to infer ecological observation suggests that, for monitoring programs, sampling to overestimated the abundance of highly susceptible taxa. This in assigning all copepods to the same category, we could have revealed copepods to have high sensitivity to OA but our analysis failed to capture variation at this level, and of copepods, not all of which are likely to be highly susceptible taxa. According to our meta-analysis, survival is among the least sensitive response categories, suggesting that changes in abundance due to differential survival could take time to grow to detectable levels. Sub-lethal effects may emerge more readily, but our data indicate that sublethal effects, if operative in these populations, were not of sufficient magnitude or duration to cause measurable declines in population abundance over the period sampled. Despite the limited temporal duration of our observations, we expected that such effects may have emerged across sites because the large spatial differences in seawater conditions that we sampled across are generally consistent year after year (Feely et al., 2010; McLaskey et al., 2016; Pelletier et al., 2018). This observation is important in the context of long-term monitoring programs that use measures of zooplankton abundance as indicators of environmental conditions. The strength of inferences made possible by monitoring data is likely to be dependent on the strength of the relationship(s) between OA conditions and the abundance of specific taxa. It is important that the scale of the response variable is adequate to reveal effects over the temporal and spatial scales sampled.

Even so, abundance is a logical first factor to examine zooplankton response to OA based on the expectation that significant effects of low pH or high pCO$_2$ on development, reproduction, and other traits that influence individual fitness ultimately will be reflected in measures of population abundance. Moreover, changes in abundance can serve as an indicator of zooplankton response to OA. For example, Smith et al. (2016) used natural gradients in CO$_2$ to show significant reductions in zooplankton biomass and abundance under high CO$_2$ conditions. Other taxa have shown similar declines in abundance in response to increasing pCO$_2$ or acidity (Hall- Spencer et al., 2008; Cigliano et al., 2010; Kroeker et al., 2011). Measures of abundance are commonly used for zooplankton observations in time series, and relative shifts in species abundances have been used extensively to examine zooplankton responses to climate (e.g., Mackas et al., 2007; Mackas and Beaugrand, 2010; Peterson et al., 2017) and as input to models of food web and ecosystem response to OA (Busch et al., 2013; Marshall et al., 2017). Abundance data can also capture shifts in a species’ center of abundance, indicating the directionality and magnitude of responses to climate change (Chivers et al., 2017). Hence, understanding the utility and limitations of abundance measures to broader environmental and ecological inquiries is
essential to interpretation of empirical studies and to natural resource management. Importantly, the results of our study suggest that the relationship between measures of zooplankton abundance and seawater pH is not simple, emphasizing the utility of intensive sampling over longer time periods to elucidate underlying relationships. Longer time series that pair chemical and biological observations and distinguish the effects of seawater carbonate chemistry from factors such as temperature and dissolved oxygen will improve attribution of biological effects among co-occurring stressors (Doo et al., 2020). Our results underscore the importance of long-term monitoring of species abundance to determine the direction and magnitude of change associated with climate and OA.

Low statistical power is common among small datasets, and interpretation of our results is further limited by the existing literature. Many zooplankton monitoring efforts lack long-term and consistent data even though it is increasingly recognized that decades-long time series will be needed to confidently identify the effects of OA on zooplankton abundance. Our findings corroborate this by demonstrating that snapshots of zooplankton abundance in relation to seawater pH can fail to detect effects of pH on abundance and community composition, even when spatial gradients in pH are relatively steep and conditions differ sharply among sampling locations.

Although abundance is more easily and frequently measured in field settings, other variables could be more informative. Measuring calcification is a clear choice because of its known sensitivity to low values of $\Omega_{ar}$ and the relative wealth of relevant laboratory studies concerning the ability of organisms to create calcium carbonate structures under OA conditions. Recent research studying in situ impacts on calcification found reduced calcification in pteropods collected in the California Current Ecosystem (Mekkes et al., 2021) and severe dissolution in pteropods in our study area (Bednaršek et al., 2020b). Importantly, some measures of calcification can be made on preserved samples, potentially alleviating constraints of some monitoring programs. However, calcification is important to only a subset of key zooplankton taxa, and even for calcifiers, is only one of many processes affected by OA. Where possible, adding biological measures in addition to abundance may provide key insights into sensitivity and add substantial value to monitoring programs. Ultimately, including mechanistic studies that can reveal cause-and-effect relationships between OA conditions and biological response will be more informative than correlation-based studies alone.

Overall, we found that observations of zooplankton abundance and community composition in response to pH and pCO$_2$ did not match experimental evidence from the literature. We propose that this mismatch could be explained by a number of factors, including response to co-occurring physical factors such as temperature and dissolved oxygen, data limitations, and biological and ecological factors, including food availability, competition and predation, local adaptation, life history stage, and the taxonomic level at which data are aggregated. Accounting for such factors in monitoring studies should improve our ability to attribute change to specific causes. Finally, and perhaps most importantly, our results underscore the importance of testing evidence from laboratory studies against the response of natural populations in defined settings. Context clearly matters, and assumptions to the contrary are likely to lead to equivocal or even misleading findings.

**DATA AVAILABILITY STATEMENT**

The raw data supporting the conclusions of this article will be made available by the authors, without undue reservation.

**AUTHOR CONTRIBUTIONS**

KK co-designed the research, performed the meta-analysis and field sampling, analyzed results, and led preparation of the manuscript. TK co-designed and supervised the research. JK contributed to the research design and supported laboratory analysis. AM supported the research cruises, collaborated on collection of the field data, and processed environmental samples. All authors contributed to preparation of the manuscript and approved the version submitted.

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**Conflict of Interest:** The authors declare that the research was conducted in the absence of any commercial or financial relationships that could be construed as a potential conflict of interest.

The handling editor declared a past co-authorship with one of the authors, AM.

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