Co-benefits and trade-offs in the water–energy nexus of irrigation modernization in China

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Abstract

There are strong interdependencies between water use in agriculture and energy consumption as water saving technologies can require increased pumping and pressurizing. The Chinese Government includes water efficiency improvement and carbon intensity reduction targets in the 12th Five-Year Plan (5YP, 2011–2015), yet the links between energy use and irrigation modernization are not always addressed in policy targets. Here we build an original model of the energy embedded in water pumping for irrigated agriculture and its related processes. The model is based on the physical processes of irrigation schemes and the implication of technological developments, comprising all processes from extraction and conveyance of water to its application in the field. The model uses data from government sources to assess policy targets for deployment of irrigation technologies, which aim to reduce water application and contribute to adaptation of Chinese agriculture to climate change. The consequences of policy targets involve co-beneficial outcomes that achieve water and energy savings, or trade-offs in which reduced water application leads to increasing greenhouse gas (GHG) emissions. We analyze irrigation efficiency and energy use in four significant provinces and nationally, using scenarios based on the targets of the 12th 5YP. At the national scale, we find that expansion of sprinklers and micro-irrigation as outlined in the 5YP would increase GHG emissions from agricultural water use, however, emissions decrease in those provinces with predominant groundwater use and planned expansion of low-pressure pipes. We show that the most costly technologies relate to trade-offs, while co-benefits are generally achieved with less expensive technologies. The investment cost per area of irrigation technology expansion does not greatly affect the outcome in terms of water, but in terms of energy the most expensive technologies are more energy-intensive and produce more emissions. The results show that water supply configuration (proportion of surface to groundwater) largely determines the potential energy savings from reductions in water application. The paper examines the importance of fertigation and highlights briefly some policy implications.

1. Introduction

China faces its own ‘perfect storm’ as rapid economic transition drives increasing per capita demand for water, energy and food. Whilst national food production has increased substantially in recent decades the agricultural sector has become responsible for nearly two-thirds of total water use and 17%–20% of China’s...
greenhouse gas (GHG) emissions (Wang et al. 2012). Agricultural water use is declining due to increasing competition for urban and industrial uses, but the agricultural sector still consumes roughly 60% of total water use (Wang et al. 2005, MWR 2014). Groundwater use in China increased from around 10 km$^3$ during the 1950s to more than 110 km$^3$ in 2010 (Shah 2009) causing increases in pumping-related emissions of GHGs (Wang et al. 2012). The interdependencies between water and energy are increasingly recognized by researchers (e.g. Hoff 2012), however, they often remain absent in planning and operation. One example is the potential for co-benefits or trade-offs to occur when attempting to reduce water application in irrigated agriculture, because energy is generally required for increased pumping, pressurizing and conveyance: situations with groundwater use for irrigation might provide co-benefits via energy savings from reduced pumping and water application (Zou et al. 2013); while situations with surface water might induce trade-offs between reductions in water application and increases in emissions when energy-intensive irrigation technology is deployed. Outcomes can be quantified and explored using a water–energy nexus perspective (WEF 2011, Sanders and Webber 2012, Hightower et al. 2013, Scott 2013, DoE 2014, Finley and Seiber 2014, Frumhoff et al. 2015, Healy et al. 2015, Isemann and Tidwell 2015). Irrigation comprises the second largest contribution (22%) to the total carbon footprint of crop production in China (Cheng et al. 2011). Jackson et al. (2010) reported emissions savings of 12%–44% could be achieved through the adoption of irrigation technology in groundwater fed conditions. Optimal management measures such as dynamic regulation of pressure may also lead to additional energy savings of 20% (Diaz et al. 2009). Irrigation technologies can reduce the non-productive consumption (Perry 2011) of water in different ways; they can minimize leaching and evaporation and improve application control to optimize water uptake in plants. Canal lining and pipelines are measures to avoid non-productive evaporation and leaching of water during conveyance; sprinkler and drip systems limit the application of water to the fields/crops. By reducing applied water and potentially helping to cope with climate change (e.g. precipitation variability and more intense droughts), irrigation technology can contribute to adaptation and, in some contexts, irrigation technology might simultaneously contribute to carbon intensity reduction and mitigation of climate change.

The Chinese Government included water efficiency, farming modernization and improved risk management in agriculture as priorities in the 12th Five-Year Plan (5YP, 2011–2015) (CPC 2011) and the No. 1 Central Document in 2011 (CPC 2010). The 12th 5YP was the first to set targets for carbon intensity reduction. Water pumping is a major input for Chinese agriculture, yet the links between energy use and the modernization of irrigation are not addressed directly in these policy plans. In fact, vertical planning processes often fail to integrate water–energy considerations (Yu 2011). The 12th 5YP targeted increases in irrigation efficiency of 3% and included objectives for increasing the area under four irrigation technologies, which are the focus of this study: sprinkler irrigation, micro/drip irrigation, canal lining for seepage control, and low pressure pipelines. The policy targets were intended to be realised by giving incentives to and supporting farmers to adopt irrigation technology through extension activities, subsidies, discount loans for equipment and water pricing (NEA 2012, Gremades et al. 2015). At the provincial level the 12th 5YP sets targets for increases in the area using irrigation technology of 26% in Hebei, 54% in Heilongjiang, 33% in Shandong and 53% in Xinjiang (CPC 2011).

This paper examines how to decrease energy consumption linked to irrigation, whilst reducing irrigation water application and maintaining food security. This objective is challenged by the diversity of situations in China, meaning that similar policies can produce either trade-offs or co-benefits. We develop an original model of the main components of the water–energy nexus of irrigation modernization and assess the consequences of sectoral policy targets in China’s 12th FYP (Scott et al. 2011). First, our assessment focuses on co-benefits, to identify win–win outcomes which achieve reductions of water applied and energy savings. Second, our assessment focusses on trade-offs, in which reductions of water applied lead to increasing GHG emissions from energy use. Third, our study assesses the possible trade-offs and co-benefits in provinces with different contexts to understand the consequences of sectoral policy goals. Fourth, our analysis considers economic aspects linked to co-benefits and trade-offs. The study uses data from government sources and develops a method to estimate changes in water use efficiency and energy use emissions. The analysis is done nationally and in four provinces with contrasting water–energy endowments (figure 1). Our assessment concludes by suggesting policy recommendations for improved management of the irrigation water–energy nexus.

2. Methodology

We develop an original model of the energy embedded in water pumping for irrigated agriculture and related processes. The model is based on the physical processes of irrigation schemes and their technological developments. The model comprises all processes, from extraction and conveyance of water to its application in the field (figure 2), and their implications in fertigation and tillage. There are many other factors that come into play in determining overall system efficiencies for the water–energy nexus. Our analysis focuses on irrigation technology—an
important (and less studied) component. The model captures changes in energy and water related to the operation of the irrigation system that are affected by technological change. The model uses data to assess water use efficiency and emissions in irrigation schemes and their technological changes. These processes may require energy (e.g. for pumping, pressurizing) and unavoidably involve some consumption of non-productive agricultural water use through evaporation, runoff and seepage.

Data inputs and calculations are based on provincial and national level data:

- Statistical Yearbook of China’s Water Resources (Ministry of Water Resources, 2011).
- Groundwater Level Yearbook of China, GEO-Environmental Monitoring Institute (year 2006).
- Extensive survey data collected by the Center for Chinese Agricultural Policy (2004–2014).
- Additional data from the Ministry of Water Resources.

Using these sources we calculate the scheme irrigation efficiency and energy use emissions for the year 2010 and use these as a baseline. The scheme irrigation efficiency, defined as the fraction of water ‘pumped or diverted through the scheme inlet which is used effectively by the plants’ (FAO 1989), is the combined efficiency of the conveyance and the application (see equation (1)). Conveyance efficiency represents the efficiency of water transport in canals and field application efficiency represents the efficiency of water application in the field

\[
\eta = \eta_c \times \eta_a, 
\]
\[
\eta_c = \sum_i A_i \eta_{c_i}, 
\]
\[
\eta_a = \sum_j B_j \eta_{a_j}, 
\]
Table 1. Pressure ranges of irrigation systems.

| Type of System          | Pressure Range (kPa) | Head (m) | Irrigation System                                      |
|-------------------------|----------------------|----------|--------------------------------------------------------|
| Low pressure            | 0–172.37             | 0–17     | Most surface and some drip/micro systems              |
| Medium pressure         | 179.26–406.79        | 18–41    | Some drip/micro and some sprinkler systems            |
| High pressure           | ≥413.69              | ≥42      | Some sprinkler systems                                 |

Source: USDA (2012).

where \( i \) is the set that includes the following delivery facilities: unlined canal, lined canal and low pressure pipeline; \( j \) is the set that includes flood, sprinkler, micro-irrigation and other kinds of irrigation technologies; \( A_j \) is the percentage of irrigated areas using the \( j \)th kind of delivery facility; \( B_j \) is the percentage of irrigated areas using the \( j \)th kind of irrigation technologies in the field; \( η_c \) is the conveyance efficiency for non-lined canal (0.30), seepage controlled canals (0.75), low pressure pipelines (0.95); \( η_a \) is the field application efficiency for flood (0.50), sprinkler (0.75), micro-irrigation (0.90) and other technologies for application (0.70), and \( η \) is the scheme irrigation efficiency.

According to the Food and Agriculture Organization of the United Nations (FAO 1989) a scheme irrigation efficiency of 50%–60% is good, 40% is reasonable, whereas 20%–30% is poor. The combined irrigation scheme efficiency depends on the type of conveyance and application system all of which have different water use efficiencies. If well maintained, long, sandy soil canals have maximum conveyance efficiencies of 60%. Taking into account poor maintenance of canals the conveyance efficiency of earthen canals is set to 30%, and the conveyance efficiency of canals with seepage control is set to 75% (FAO 1989). For low pressure pipelines, we assume a conveyance efficiency of 90%. For application efficiencies we adopt a mid-value of 50% for flood systems, 75% for sprinkler, 90% for drip and 70% for other field practices that enhance efficiency.

To calculate the scheme irrigation efficiencies (equation (1)), we use national statistics for irrigated areas by province. For conveyance efficiency, the total irrigated area is divided according to the areas supplied by seepage controlled canals, low pressure pipelines and the remaining area is assumed to be supplied by traditional non-lined canals (see equation (2)). For field application efficiency, the total irrigated area is divided according to areas of sprinkler, micro and other field measures; the remaining is classified as flood (see equation (3)). Flood irrigation comprises all types of surface water irrigation (basin, furrow, border). It is important to note that Chinese government metrics for irrigation efficiency do not consider conveyance and application practices separately. While areas supplied by water through lined canals and pipelines are counted as irrigation technologies, water might still be applied through flooding practices. Hence, we argue it is important to acknowledge that water can leave the irrigation scheme through non-productive evaporation and seepage both during conveyance and application and therefore both steps should be incorporated in any assessment.

For the energy use emissions, we divide the energy used for an irrigation scheme into pumping and conveyance (equation (5)), and pressurizing water for application to the fields (equation (6)). For both calculations we consider the two main types of engines used for water pumping in China: according to Wang et al (2012) the proportion of engines used for water pumping in China: according to Wang et al (2012) the proportion of engines used is set to 76% for electric and 24% for diesel. Similarly, pump engine efficiencies of electric and diesel driven systems are set to 0.40 and 0.15, respectively, and the efficiency of the electric system incorporates transmission and distribution losses assumed at 15% (Wang et al 2012). Derating efficiency is considered and set at 80% for electric driven systems and 75% for diesel. Pump head is the most crucial factor, noting that efficiency of the power generation and supply and the pump and pipeline system also influence energy use. To calculate the energy used in the pumping and conveyance of groundwater and surface water we use equation (5) (Rothausen and Conway 2011). Conveyance consumes energy in the case of low pressure pipes. Even so, results from a large-scale field survey (see Cre- mades et al 2015) show low pressure pipes are used in 91% of cases in areas irrigated exclusively with groundwater, only 2% of cases in areas irrigated exclusively with surface water, and the remaining 7% in areas with a mix of both. We therefore assume that the expansion of low pressure pipelines will occur in areas irrigated exclusively with groundwater. Since the adoption of low pressure pipelines implies lower extraction of groundwater, the final effect is less water pumped and less energy consumed, as found by Zou et al (2015).

We use equation (4) to calculate the groundwater pump heads for each of the provinces, according to regression estimates from survey data in Wang et al (2012) relating groundwater levels with pump heads

\[
\text{Pump head} = 0.906 \times \text{groundwater level} + 21.75.
\]

The energy necessary for groundwater pumping and conveyance, and resulting emissions, is calculated using equation (5). The characteristics of the energy
supply in the provinces are taken into account in the calculation of groundwater pumping.

\[
\text{Emissions from groundwater pumping and conveyance (kgCO}_2\text{e)} = \sum_{\text{en},m,r} \left[ \frac{9.8(\text{m s}^{-2}) \times \text{Pump head}_{en,r}(\text{m}) \times \text{Water pumped}_{en,r}(\text{kg})}{3.6 \times 10^6 \times \text{Derating}_{en,\%} \times \text{Pump efficiency}_{en,\%}} \times \text{Conversion factor}_{en,r}\left(\text{kgCO}_2\text{e kWh}^{-1}\right)} - A_r(\text{ha}) \times LP\left(\text{kgCO}_2\text{e ha}^{-1}\right) \right].
\] (5)

where \(k\) is the set of energy intensive water sources, comprising pumping of groundwater and pumping of surface water; \(m\) is the set including both engine types, diesel and electric; \(r\) is the set that includes the regions studied, i.e. the provinces of China; emissions from groundwater pumping and conveyance are the comprising emissions caused by energy-intensive water sources and decrease of emissions related to conveyance using low pressure pipes, the calculation is analogous for each province and the result is shown at country level, summing over all the provinces; pump head: also referred as lift or pump lift, vertical distance over which water is lifted, or the combined effect on pressure of elevation and/or distance over which water is pumped prior to application, in groundwater extraction the pump head is larger than the groundwater level, because a lowering is created in the level of the aquifer, pump head varies for different water sources (groundwater and surface water) and provinces; water pumped is the the weight of the water pumped, its amount varies for different water sources (groundwater and surface water) and provinces; pump efficiency is the performance of the pump in terms of conversion to mechanical energy; conversion factor: as per the calculation in equation (6) below, which varies for different provinces and engines; AD is the area with conveyance through low pressure pipes, which varies for different provinces; \(D\) is the reduction of emissions due to changes in ploughing and tilling operations after the adoption of drips, 20 kg of diesel per ha as reported by Li et al (2016).

In equation (7) we create conversion factors to derive the emissions per energy use for diesel and electricity produced in China with a provincial detail

\[
\text{Conversion factor}_{en,r}(\text{kgCO}_2\text{e kWh}^{-1}) = \text{Engine}_{en} \times \text{Emission factor}_{en,r}(\text{kgCO}_2\text{e kWh}^{-1}),
\] (7)

where conversion factor is the emissions for the two main types of energy use for water pumping in China, diesel and electricity, engine is the distribution of
engines used, 76% for electric and 24% for diesel, as detailed above, emission factor is the emissions for electricity produced in China with a detail on the provinces (Lindner et al. 2013) and conversion factors for diesel (DEFRA and DECC 2010).

The nexus between water use and emissions is captured for pressurized application and for pressurized conveyance. First, the emissions caused by irrigation in areas adopting pressurized conveyance are represented in equation (4); when these areas vary, the change is also reflected in equations (2) and (1). Second, the emissions caused by irrigation in areas adopting pressurized application technologies are represented in equation (5); when these areas change, it is also reflected in equations (3) and (1). Therefore, the method captures the implications in terms of both irrigation efficiency and emissions.

Using this procedure, we calculate a baseline and simulate the water and energy use associated with increasing the area that adopts irrigation technology as described in the 12th 5YP targets. Then, by distributing the reductions of water applied from the estimated increases in irrigation efficiency, according to the supply sources in each province, we evaluate the potential GHG emission savings. Finally, we quantify the influence of key parameters in the formulae above to understand their impact on irrigation efficiency and energy use. Figure 2 shows the main data sets used in the methodology.

This approach is used to analyze the outcome of the 12th 5YP targets in four Chinese provinces: Hebei, Heilongjiang, Shandong, and Xinjiang.
Heilongjiang, Shandong and Xinjiang (figure 1). The provinces have been chosen because they had the highest planned increases in area under irrigation technology according to the 12th 5YP. The four selected Chinese provinces differ substantially in key parameters like agricultural water use, sources of water, irrigated area, area under irrigation technology, and conveyance and or application type (more information in supplementary section 1). These differences in key variables lead to very different values of fundamental aspects of the water–energy nexus such as energy emission rate, total emissions, irrigation efficiency, water savings and the costs of the related investments.

3. Results and discussion

3.1. Baseline situation

Overall, GHG emissions from China’s irrigation for a baseline scenario of 2010 are estimated to be 37.66 Mt CO₂e (table 2). This result comprises energy use from groundwater pumping and pumped surface water, including conveyance and application in irrigation systems. This result is similar to previous findings by Zou et al (2015), however, our estimate for groundwater pumping on its own (22.37 Mt CO₂e) is significantly lower than previous estimates of 28.65 Mt CO₂e by Zou et al (2015), and of 33.1 Mt CO₂e by Wang et al (2012). The contrasting results are likely to be due to the cumulative effects of minor differences in methods between the studies, and perhaps more importantly, due to our inclusion of updated data at the provincial level from the energy sector, which is rapidly evolving.

The irrigation efficiency value calculated for the baseline is 23.80% (table 2), which is lower than the FAO estimate for East Asia (33%; FAO 2002), and much lower than the Chinese official irrigation efficiency figure for 2010 (50%; MWR 2011). It should be noted that estimates of irrigation efficiency are notoriously uncertain at all scales and they are subject to strong debate. Differences in the definition and/or calculation method can greatly influence results. Our definition of irrigation efficiency is the fraction of water ‘pumped or diverted through the scheme inlet which is used effectively by the plants’ (FAO 1989), while the Chinese government definition is the ratio between water available for crops and water extracted (Han et al 2009, MWR 2009). The definition we use includes conveyance efficiency, which is absent in the official Chinese definition (i.e. we include water abstracted from source to crop, whereas the government definition only includes water from irrigation scheme source to crop). Hence, the two irrigation efficiencies are not directly comparable.

The socio-ecological contexts of the four provinces are reflected in contrasting relative distributions of water supply sources and adoption levels of different irrigation technologies and therefore show considerable differences in both the overall efficiency and between conveyance and application types (table 2). Hebei has a very efficient conveyance system partly attributed to widespread use of low pressure pipelines but predominantly to the high proportion of groundwater use for irrigation (77%). CO₂e emission rates show even larger differences as rates vary considerably between groundwater and surface water pumping, depending on the pump head. Hebei has high total emissions from irrigation because of its large proportion of groundwater use. The relatively extensive use of sprinkler systems in Heilongjiang leads to higher emissions than other provinces, whereas the greater pump head in Xinjiang causes its high emissions.

3.2. Scenario results

We next project the effects of expanding the area under irrigation technology according to the 12th 5YP plan.

3.2.1. Irrigation efficiency and reductions in water application

Table 3 shows an increase in irrigation efficiency from 23.80% to 26.76% which is consistent with the 3% increase targeted by the national 12th 5YP. Hebei and Heilongjiang show higher increases in irrigation efficiency than the target (Hebei 3%, Heilongjiang 3%, Shandong 4%, and Xinjiang 6%). If realised, these changes would provide reductions of applied water of nearly 11 km³ that could be used to intensify or expand irrigated cropping to help meet the 5YP target of increased grain production. The national average reduction in applied water is 170.93 m³ ha⁻¹ yr⁻¹ and the studied provinces range from 139.40 to 502.09 m³ ha⁻¹ yr⁻¹. Hebei, with the highest irrigation efficiency, has the lowest reduction in water application, whereas Xinjiang, with an irrigation efficiency slightly above the national value, has the highest reduction in water application. These differences highlight the importance of increasing efficiency in areas with high water use and low efficiency.

3.2.2. Emissions from fuel consumption

The expansion of pressurized irrigation systems leads: first, to increases in water pressurization for sprinklers and drips, which consume energy that comes from different provincial mixes; second to changes in farming operations, namely a decrease in tillage requirements in areas with drips; and third to reductions in emissions due to smaller water conveyance losses resulting in smaller water pumping needs with the use of low pressure pipes in groundwater contexts. Hence, energy savings directly related to pressurization by irrigation technology are primarily due to reductions in water use, so pressurization-related energy is saved chiefly if less water is pumped and/or pressurized. These aspects cause an increase in the GHG emissions of 0.26% (0.10 Mt CO₂e) for China as a whole. Table 3
shows that the effects on emissions from energy use for pressurization vary substantially among the provinces. Heilongjiang experiences increases of around 25%; Hebei and Shandong, on the other hand, show small decreases due to their adoption of low pressure pipes and proportionally lower increases in sprinkler and micro-irrigation systems.

3.2.3. The economics of the water–energy nexus of irrigation technology

A review of the investment costs of different irrigation technologies in 13 Chinese studies (see supplementary table S6) allows an estimate of average cost (CNY ha⁻¹) for each of the four irrigation technologies that can be used to calculate the reduction in water application per unit cost. Improving conveyance by canal lining costs roughly 4300 CNY ha⁻¹, low pressure pipes cost roughly 3750 CNY ha⁻¹, sprinklers 9700 CNY ha⁻¹ and micro-irrigation is the most expensive at 14 500 CNY ha⁻¹. Figure 3 shows varying investment costs per area in different provinces, according to different areas adopting the technologies. In the same figure, when investment per area is compared with the change in water use per area across the provinces, they

| Table 3. Increase in the area adopting irrigation technology as targeted in the 12th 5YP for 2015 and the effects on scheme irrigation efficiency and GHG emissions. |
|---------------------------------------------------------------|
| **Units** | **National** | **Hebei** | **Heilongjiang** | **Shandong** | **Xinjiang** |
|---------------------------------------------------------------|
| **Targeted expansion of irrigation technology** | | | | | |
| Sprinkler | 1000 ha | 2001 | 19 | 1115 | 33 | 90 |
| Microirrigation | 1000 ha | 2011 | 19 | 87 | 33 | 1297 |
| Low pressure pipes | 1000 ha | 2453 | 559 | 1 | 522 | 27 |
| Canal lining | 1000 ha | 3535 | 108 | 242 | 162 | 172 |
| Total | 1000 ha | 10000 | 705 | 1445 | 750 | 1586 |
| **Irrigation efficiency achieved with the targeted areas** | | | | | |
| **under advanced irrigation technology** | | | | | |
| Conveyance | % | 48.37 | 66.98 | 34.39 | 55.63 | 42.38 |
| Application | % | 55.32 | 52.46 | 73.03 | 52.69 | 72.56 |
| TOTAL | % | 26.76 | 35.14 | 25.11 | 29.31 | 30.75 |
| % Total increase | % | 12.45 | 14.67 | 22.53 | 16.27 | 21.11 |
| Absolute figure of total increase | % | 2.96 | 4.49 | 4.62 | 4.10 | 5.36 |
| **Change in emissions from energy consumption** | | | | | |
| Groundwater pumping | MtCO₂ yr⁻¹ | 22.37 | 6.83 | 4.30 | 2.38 | 2.92 |
| Surface water pumping | MtCO₂ yr⁻¹ | 11.75 | 0.11 | 0.51 | 0.56 | 0.14 |
| Pressurizing water for conveyance (low pressure pipes) | MtCO₂ yr⁻¹ | −1.62 | −0.45 | 0.00 | −0.30 | −0.02 |
| Pressurizing water for application (sprinkler) | MtCO₂ yr⁻¹ | 4.99 | 0.18 | 2.93 | 0.16 | 0.29 |
| Pressurizing water for application (micro irrigation) | MtCO₂ yr⁻¹ | 1.58 | 0.01 | 0.12 | 0.03 | 1.76 |
| Decrease in tillage operations (micro irrigation) | MtCO₂ yr⁻¹ | −0.26 | 0.00 | −0.01 | −0.01 | −0.18 |
| TOTAL | MtCO₂ yr⁻¹ | 38.11 | 6.68 | 7.84 | 2.83 | 4.91 |
| % Total variation from baseline | % | 3.07 | −1.21 | 26.61 | −1.84 | 20.80 |
| Absolute figure of variation from baseline | % | 1.16 | −0.08 | 1.65 | −0.05 | 0.85 |
| **Change in emissions from fertigation** | | | | | |
| Reduction in emissions due to smaller nitrogen application | MtCO₂ yr⁻¹ | −1.06 | −0.14 | −0.15 | −0.13 | −0.25 |
| **Overall effects of irrigation technology on emissions, including fertigation, compared with baseline** | | | | | |
| TOTAL | MtCO₂ yr⁻¹ | 37.75 | 6.54 | 7.70 | 2.70 | 4.66 |
| % Total variation from baseline | % | 0.26 | −3.23 | 24.23 | −6.35 | 14.77 |
| Absolute figure of variation from baseline | MtCO₂ yr⁻¹ | 0.10 | −0.22 | 1.50 | −0.18 | 0.60 |
| **Overall effects of irrigation technology on emissions, assuming technology is used consuming same amounts of water and fertilizer, compared with baseline** | | | | | |
| TOTAL | MtCO₂ yr⁻¹ | 44.71 | 6.68 | 7.84 | 2.83 | 4.91 |
| % Total variation from baseline | % | 18.72 | −1.21 | 26.61 | −1.84 | 20.80 |
| Absolute figure of variation from baseline | MtCO₂ yr⁻¹ | 7.05 | −0.08 | 1.65 | −0.05 | 0.85 |
visibly follow a similar proportion. Strikingly, the actual reduction in water application per unit of investment remains very similar across the case studies (Figure 3): in the studied provinces, the reduction in water application per unit of investment varies from 0.33 (Hebei) to 0.38 (Xinjiang) hectolitre yr⁻¹ CNY⁻¹. That is to say, that in the studied provinces the output has a similar proportion to the investment whatever technological change is made. Importantly, the relative investment returns in terms of water remain similar, but different technological changes create different outcomes in terms of emissions, which are strongly related to the level of expenditure, because lining canals is cheaper than installing drips or sprinklers. Summarizing, the investment cost per area does not have large implications for the outcome in terms of water, but in terms of energy the most expensive technologies listed above are more energy-intensive and therefore produce more emissions.

3.2.4. Regional details strongly influence outcomes

The characteristics of the water supply systems strongly determine the potential for energy savings from reductions in water application. Hebei and Shandong show modest co-benefits, with lower emissions compared to the baseline, which can be explained by the characteristics of their water supply system. A high proportion of groundwater and deep groundwater pump lift make Hebei’s water supply very energy intensive, so that reductions in applied water tend to reduce emissions. This finding is consistent with the potential for energy savings in Hebei presented by Zhang et al (2013).

In Shandong the increase in areas adopting low pressure pipes offsets the increase of emissions due to larger areas adopting micro-irrigation and sprinklers. Both Heilongjiang and Xinjiang exhibit increases in emissions. For instance, despite the relatively large reductions in water application in Xinjiang, emissions show a significant trade-off, increasing by 15%. Co-benefits are lacking as the main water supply in Xinjiang is gravity-fed surface water. In terms of GHG emissions, pressurized irrigation technology such as sprinklers and micro-irrigation is best implemented in areas with energy-intensive water supply. Where this is not the case, it is important to recognize the energy trade-offs associated with implementing energy-intensive irrigation technology, as seen in the case of Heilongjiang where emissions increase nearly 25%, despite reductions in water application.

Prioritizing low pressure pipes in contexts dominated by groundwater sources, like Hebei, is cheaper and leads to co-benefits. A similar result can be expected from canal lining; although it is not a technology strongly targeted in the analyzed provinces, it could play a significant role. In contrast, prioritizing more expensive technologies like sprinklers and micro-irrigation leads to costly trade-offs, due to increased energy consumption when using surface water and/or increasing the pressure for sprinkler devices.

3.2.5. The potential influence of reduced emission from fertigation

The changes in irrigation systems that we describe have implications for the amount of nitrogen fertilizer used (a key contributor to overall agricultural emissions). Fertilizer is applied jointly with irrigation in the pressurized schemes, which improves the efficiency of the application, and a reduction of nitrogen application would be expected. This additional trade-off is not included in the baseline because there is no detailed data on the actual amount of nitrogen fertilizer used in different provinces and management systems. Moreover there is no strong empirical evidence to estimate the reduction of fertilizer use that would be implemented by Chinese farmers (who are known to generally use much more nitrogen than needed, Qiu 2009). Indeed, the adoption areas of
sprinklers and drips in the field in China are very limited and many are still in pilot project sites. In the literature there are estimates of ranges of nitrogen application values for different crops (Yang et al. 2015), but they lie below the values shown by the few sources that do provide survey based data for such technologies (Wang et al. 2008). Given this uncertainty about how fertigation will be managed and with a note of caution, we provide an estimate (table 3) by assuming a 30% reduction as proposed in government guidelines for fertigation (MA 2013), taking as departure point the application values reported by Yang et al. (2015) for those areas where pressurized irrigation technology is adopted. Following IPCC guidelines, the emission factors used are country specific (Zhou et al. 2015), according to Tier 2 methodology and differentiated between rice and cropland (Stocker 2014), and the values are converted to CO2 equivalent using global warming potential values for 100 years (Stocker 2014). Importantly, in cases where there is a large expansion of micro-irrigation, like in Xinjiang (table 3), the decrease of emissions due to reduced application of nitrogen fertilizer makes a visible difference, but it does not change the overall picture.

3.2.6. Perverse outcomes: maximizing production
The relative advantage created by the adoption of drips and sprinklers and the related decrease in marginal costs of water can create an unexpected outcome, namely that drips and sprinklers are used for maximizing production rather than for decreasing water application; an outcome that is being documented in, for example Pakistan and the US (Ahmad et al. 2007, Ward and Pulido-Velazquez 2008); water use actually increases after technology adoption creates a rebound effect. This is likely in situations of weak water governance structures and implementation, with neither effective control of water use nor economic mechanisms to regulate overuse, conditions typical in much of China. A scenario in which technology improvements are associated with the same consumption of water and fertilizer shows that co-benefits would still appear in the provinces with less drip and sprinkler technology, but for the whole of China it would mean a significant increase of emissions (table 3, last row).

3.3. Influence of key parameters on the water–energy nexus of irrigation modernization
Finally we examine the influence of incremental changes in some key parameters at the national scale. Increases in the area of canal lining for conveyance and micro-irrigation for application produce the highest proportional effect on raising irrigation efficiency (figure 4). The analysis of the influence on emissions from irrigation technology, and water and energy sources, shows that expansion of sprinkler systems strongly increases the emission rates (figure 5); this would not be the case, however, in areas where the groundwater pump head would be greater than the critical energy saving head (Zou et al. 2012). There are other important messages: pumping groundwater is clearly the most energy intensive process and changing the distribution of energy sources towards more electric pumps produces a reduction in emissions.

The predominant use of surface water and traditional unlined canals is the main reason for the relatively low irrigation efficiency in China. Consequently, improving conveyance efficiency through canal lining and pipelines may be the most effective way of achieving the co-benefit of reducing water application with a low carbon footprint. Only in groundwater fed areas is micro-irrigation a suitable irrigation technology for both improving irrigation efficiency and decreasing GHG emissions, in situations where the pumping head value allows it (Zou et al. 2012).

4. Conclusions and policy implications
Quantifying the trade-offs and co-benefits is just part of the nexus story; to successfully manage the water–energy nexus in China, a critical set of socio-economic and policy issues also need to be addressed. These include: improving communication amongst ministries

Figure 4. Influence of incremental percent changes in the area using irrigation technology on irrigation efficiency.

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**Table 3**

| Irrigation Technology | Emission Reduction (%) |
|-----------------------|------------------------|
| Canals                | 26.7                   |
| Pipelines             | 27.2                   |
| Sprinklers            | 27.5                   |
| Micro                 | 27.6                   |

**Figure 4**

Influence of incremental percent changes in the area using irrigation technology on irrigation efficiency.
with responsibility for different aspects of the water–energy nexus, creating clear incentives to decrease water application and adopt irrigation technology, establishing reliable water rights, and improving groundwater governance and use (Zhang 2007, Zhang et al 2008, Calow et al 2009, Shiferaw et al 2009, Yu 2011). This study enhances our understanding of co-benefits and trade-offs in the irrigation water–energy nexus, across a set of diverse provincial cases in China using targets defined in the 12th 5YP. The approach applies a suite of assumptions about water use efficiency for various stages and technologies in the process of irrigation at the provincial level, although estimates for the energetic implications of low pressure pipes were only available on a per area basis at the national scale.

We find that expansion of sprinklers and micro-irrigation as outlined in the 12th 5YP could increase GHG emissions from agricultural water use by just 0.10 MtCO2-e yr−1, however, under the existing weak governance this could increase to 7.05 MtCO2-e yr−1 if unplanned rebound effects occur. Where pressurized irrigation technology is used in surface water irrigated areas (Xinjiang) or where the energy consumption to pressurize water (Heilongjiang) is high, emissions increase substantially. The results show that water supply configuration largely determines the potential energy savings from reductions in water application. An implication of this is that co-benefits of irrigation technology for energy saving only appear in areas irrigated with energy intensive supply (Hebei, Shandong). Trade-offs appear in surface water irrigated areas (Xinjiang), where emissions due to expansion plans for pressurized irrigation technology could increase significantly. Taken together, these results suggest that in situations where policy makers seek to optimize both water and energy use they should encourage the adoption of low pressure conveyance pipes in groundwater irrigated areas, and canal lining in surface water areas, since these increase efficiency with lower emissions than other methods. Another important implication is that sprinklers and micro-irrigation appear as a suitable means to increase efficiency only in groundwater irrigated areas. Regarding the costs, the most expensive technologies appear linked to trade-offs, while co-benefits result from lower-priced technologies. These insights are relevant to make adaptation policies consistent with mitigation goals in the irrigation sector. Further work needs to be done at the sub-provincial scale, including collection of data at smaller administrative units, and examining in greater detail which governance schemes are needed to achieve co-beneficial outcomes in different contexts.

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References

Ahmad M, Giordano M, Turral H, Masih I and Moosd Z 2007 At what scale does water saving really save water? Lessons from the use of resource conservation technologies in Pakistan J. Soil Water Conserv. 62 29–35
Calow R C, Howarth S E and Wang J 2009 Irrigation development and water rights reform in China Int. J. Water. Resour. 25 227–48
Cheng K, Pan G, Smith P, Luo T, Li L, Zheng J, Zhang X, Han X and Yan M 2011 Carbon footprint of China’s crop production— an estimation using agro-statistics data over 1993–2007 Agric. Ecosyst. Environ. 142 231–7
CPC 2010 Number 1 Document for 2011: Decision from the CPC Central Committee and the State Council on Accelerating Water
Conservancy Reform and Development (in Chinese) (http://gov.cn/gongbao/content/2011/content_1803158.htm) (accessed 5 August 2011)

CPC (Communist Party of China) 2011 12th Five-Year Plan (in Chinese) (http://gov.cn/2011lh/content/1825838.htm)

Cremades R, Wang J and Morris J 2015 Policies, economic incentives and the adoption of modern irrigation technologies in China: a review. China Earth Syst. Dyn. 6 399–410

DEFRA (Department of Environment, Food and Rural Affairs, United Kingdom) and DECC (Department of Energy and Climate Change, United Kingdom) 2010 Guidelines to DEFRA / DECC’s GHG Conversion Factors for Company Reporting (http://archive.defra.gov.uk/environment/business/reporting/pdf/100805-guidelines-ghg-conversion-factors.pdf)

DoE (United States Department of Energy) 2014 The Water–Energy Nexus: Challenges and Opportunities (Washington: United States Department of Energy)

Diaz J A R, Luque R L, Cobo M T C, Montesinos P and Poyato E C 2009 Exploring energy saving scenarios for on-demand pressurised irrigation networks Biosystems Eng. 104 532–61

FAO 1989 Irrigation Water Management: Irrigation Scheduling (http://fas.org/docrep/hto20e/hto20e88.htm)

FAO (Food and Agriculture Organization of the United Nations) 2002 Crops and Drops, Making the Best use of Water for Agriculture (ftp://ftp.fao.org/docrep/005/y3918e00.pdf)

Finley J W and Seiber J N 2014 The nexus of electricity, water and climate change Environmental Res. Lett. 10 080201

Han Z, Pei Y, Li Y and Guo H 2009 Estimation and analysis of effective utilization of irrigated water coefficient China Water Resources Res. 3 1–4

Healy R W, Alley W M, Engle M A, McMahon P B and Bales J D 2015 The Water–Energy–Nexus: an Earth Science Perspective (Reston, Virginia: United States Geological Survey)

Hightower M, Reible D and Webber M 2013 Workshop report: Understanding the nexus Earth Syst. Dyn. 4 385–395

Hoff H 2012 Understanding the nexus Background Paper for the Bonn 2011 Conf.: The Water, Energy and Food Security Nexus (Stockholm: Stockholm Environment Institute)

Ismen T and Tidwell V C 2015 Integrated modelling of the energy–water nexus in the american west Climate, Energy and Water: Managing Trade-offs, Seizing Opportunities ed J Pittoc et al (Cambridge: Cambridge University Press)

Jackson T M, Khan S and Hafeez M 2010 A comparative analysis of water application and energy consumption at the irrigated field level Agric. Water Manag. 97 1477–85

Li H, He J, Bharucha Z P, Lai R and Pretty J 2016 Improving China’s food and environmental security with conservation agriculture Int. J. Agric. Sustain. 3 1–12

Lindner S, Liu Z, Guan D, Geng Y and Li X 2013 CO2 emissions from China’s power sector at the provincial level: consumption versus production perspectives Renew. Sust. Energy Rev. 19 164–72

MA (Ministry of Agriculture of the People’s Republic of China) 2013 The Technical Guidance of Fertilization (in Chinese) (http://gov.cn/gddt/2013-03-04/content_2344628.htm)

MWR (Ministry of water resources of the People’s Republic of China) 2009 Technical Guidelines for Calculating National Irrigation Water Use Efficiency (http://jgjc.com.cn/index/Display.asp?NewId=11639)

MWR 2011 Water Resources Bulletin 2010 (Beijing: China Water Power Press)

MWR 2014 Water Resources Bulletin 2013 (Beijing: China Water Power Press)

NEA (National Energy Administration of China) 2012 Water Conservation Development plan (2011–2015) (http://nea.gov.cn/2012-06/26/c_1116773311.htm)

Perry C 2011 Accounting for water use: terminology and implications for saving water and increasing production Agric. Water Manag. 98 1840–6

Qi J 2009 Nitrogen fertilizer warning for china Nature 10 103–5

Rothausen S G A and Conway D 2011 Greenhouse-gas emissions from energy use in the water sector Nat. Clim. Change 1 210–14

Sanders T K and Webster M E 2012 Evaluating the energy consumed for water use in the united states Environ. Res. Lett. 7 3

Scott C A 2013 Electricity for groundwater use: constraints and opportunities for adaptive response to climate change Environ. Res. Lett. 8 035005

Scott C A, Pierce S A, Pasqualetti M J, Jones A L, Montz B E and Hoover J H 2011 Policy and institutional dimensions of the water–energy nexus Energy Policy 39 6622–30

Shah T 2009 Climate change and groundwater: India’s opportunities for mitigation and adaptation Environ. Res. Lett. 4 045005

Shiferaw B A, Reddy R V and Okello J 2009 Adoption and adaptation of natural resource management innovations in smallholder agriculture: reflections on key lessons and best practices Environ. Dev. Sustain. 11 601–19

Stocker T F 2014 Climate Change 2013: The Physical Science Basis: Working Group I Contribution to The Fifth Assessment Report of the Intergovernmental Panel on Climate Change (Cambridge: Cambridge University Press)

USDA (United States Department of Agriculture) 2012 Energy Estimator: Irrigation Tool (http://ipat.sc.egov.usda.gov/Help.aspx)

Wang H J, Huang B, Shi X, Darilek J L, Yu D, Sun W, Zhao Y C, Chang Q and Öborn I 2008 Major nutrient balances in small-scale vegetable farming systems in peri-urban areas in China Nutr. Cycl. Agroecosys. 80 173–83

Wang J, Rothausen S G A, Conway D, Zhang L, Xiong W, Holman I P and Li Y 2012 China’s water–energy nexus: greenhouse-gas emissions from groundwater use for agriculture Environ. Res. Lett. 7 014035

Wang J, Xie Z, Huang J and Rozelle S 2005 Incentives in water agriculture Water Policy 3 1–306

Ward F A and Pulido-Velazquez M 2008 Water conservation in irrigation can increase water use Proc. Natl Acad. Sci. USA 105 18215–20

WEF (World Economic Forum) 2011 Water Security: The Water–Food–Energy–Climate Nexus (Washington DC: Island Press)

Yang F, Meng Y, Jiang Y, Cui Y, Li O, Dong Y and Sun Z 2015 Chemical fertilizer application and supply in crop farming in China: a technical and economic analysis (Cambridge: Cambridge University Press)

Yu C 2011 China’s water crisis needs more than words Nature 470 307–12

Zhang J 2007 Barriers to water markets in the Heihe River basin in northwest China Agric. Water Manage. 87 32–40

Zhang L, Wang J, Huang J and Rozelle S 2008 Development of groundwater markets in China: a glimpse into progress to date World Dev. 36 706–26

Zhang Q, Xiao Q, Liu C K C and Geng S 2013 Technologies for efficient use of irrigation water and energy in China J. Integr. Agr. 11 1363–70

Zhou F et al 2015 New model for capturing the variations of fertilizer-induced emission factors of N2O Glob. Biogeochem. Cycles 29 885–97

Zou X, Li Y, Cremades R, Gao Q, Wan Y and Qin X 2013 Cost-effectiveness analysis of water-saving irrigation technologies based on climate change response: a case study of China Agric. Water Manage. 129 9–20

Zou X, Li Y, Gao Q and Wan Y 2012 How water saving irrigation contributes to climate change resilience—a case study of practices in China Mitig. Adapt. Strateg. Glob. Chang. 17 111–32

Zou X, Li Y, Li K, Cremades R, Gao Q, Wan Y and Qin X 2015 Greenhouse gas emissions from agricultural irrigation in China Mitig. Adapt. Strateg. Glob. Chang. 135 1–21