Short duration overnight cattle kraaling in natural rangelands: Does time after kraal use affect their utilization by wildlife and above ground grass parameters?

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Abstract

In east and southern Africa some private ranch owners are corralling (hereafter kraaling) cattle overnight for short periods (for example, seven days) in natural rangelands to create nutrient enriched hotspots which are attractive to large herbivores. However, the effect of season and time after kraal use (alt. age of nutrient enriched hotspots) on large herbivore use of these sites has not been examined. We collated the number of large herbivore sightings per day from camera traps during wet, early and late dry season in nutrient enriched hotspots of varying ages (1, 2, 3 and 4 years) and surrounding vegetation. In addition, above ground grass biomass and height in nutrient enriched hotspots was compared to that of the surrounding vegetation. Furthermore, we tested if repeated grazing in nutrient enriched hotspots stimulated grass compensatory growth. Large herbivore use of nutrient enriched hotspots was similar during wet, early and late dry season. Time after kraal use had a significant effect on mixed feeders (impala and African savanna elephant) utilization of nutrient enriched hotspots of varying ages (1, 2, 3 and 4 years) and surrounding vegetation. In addition, above ground grass biomass and height in nutrient enriched hotspots was compared to that of the surrounding vegetation. Furthermore, we tested if repeated grazing in nutrient enriched hotspots stimulated grass compensatory growth. Large herbivore use of nutrient enriched hotspots was similar during wet, early and late dry season. Time after kraal use had a significant effect on mixed feeders (impala and African savanna elephant) utilization of nutrient enriched hotspots but not grazers (zebra and warthog) and browsers (giraffe and greater kudu). Both impala and African savanna elephants mostly used nutrient enriched hotspots one year after kraal use. Aboveground grass biomass and height were higher in surrounding vegetation than in nutrient enriched hotspots. Repeated clipping (proxy for grazing) resulted in compensatory aboveground grass biomass in nutrient enriched hotspots, which declined with time after kraal use. We concluded that nutrient enriched hotspots created through short duration overnight kraaling were important foraging sites for large herbivores.

Introduction

In African savanna ecosystems availability of nutritive forage is important for both domestic and wild herbivores. Old cattle bomas or corrals (also referred to as glades) are considered
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Above ground grass biomass and height in nutrient enriched hotspots is regulated by two factors, viz. soil nutrient content and large herbivore grazing. Grass responds to improved soil fertility through rapid growth. However, attraction of grazers to nutrient enriched hotspots, in
response to availability of abundant and nutritive grass [21,22], is expected to result in intense grazing, leading to reduced aboveground grass biomass and height. The highest above ground grass biomass in nutrient enriched hotspots is recorded just after the first rains post kraaling as soil nutrients are at their peak, and thereafter decline with time after kraal use [2]. However, grazing intensity also regulates above ground grass biomass and height. In order to determine the effect of age of nutrient enriched hotspots on above ground grass biomass and height the amount of grass cropped by large herbivores needs to be measured. Impalas select short, low fiber grass which is highly digestible [22], while zebras select foraging patches with high aboveground grass biomass to achieve high digestive fill, because as hindgut fermenters they have a fast digesta passage rate [23]. However, zebra and other equids consume both short and tall grass to balance between forage quality and quantity [24]. For example, zebra have narrow muzzle considered well suited for clipping tall grasses [25]. Grass height influences herbivore habitat use [26].

Repeated grazing creates a positive herbivore-grass feedback loop that maintains high plant nutrient levels [27]. Grass resprouts are rich in nutrients and result in repeated grazing [3,28]. Nutrient enriched hotspots are subjected to repeated grazing which in the long term could result in grass failing to compensate lost biomass [29]. The ability of grass to compensate lost biomass under repeated grazing needs to be determined to better understand the effect of creating nutrient enriched hotspots in rangelands. Although repeated grazing is stimulatory to growth [30], grass responds through under-, partial- or over-compensation of lost foliar tissue [28]. Generally, most grass either under or equally- compensate the lost biomass because grazing results in loss of photosynthetic material, limiting the ability of the plants to photosynthesize [31]. Grass regrowth in nutrient enriched hotspots benefit from enhanced soil fertility due to dung and urine deposition [32,33]. In this study we investigated if repeated grazing resulted in compensatory aboveground grass biomass in nutrient enriched hotspots of different ages.

The use of nutrient enriched hotspots by large herbivores can be studied using camera traps as they are cost-effective, efficient and non-intrusive [12–14]. Camera traps can be used to determine spatial and temporal use of foraging resources by large herbivores [16]. For example, Young et al. [32] used camera traps to study forage selection by grazers in Kruger National Park. The use of foraging resources by large herbivores can be influenced by their abundance. Hence, large herbivore abundance need to be related to the number of sightings in camera traps to calculate relative abundance indices to ascertain if the use of nutrient enriched hotspots is affected by population. The probability of sighting animals in camera traps is strongly influenced by their population [29].

We studied the use of nutrient enriched hotspots of varying ages (1, 2, 3 and 4 years after kraal use) by six large herbivores of different sizes (zebra, warthog, impala, African savanna elephants, giraffe and greater kudu) and feeding guilds (grazers–zebra and warthog; mixed feeders–impala and African savanna elephant; and browsers–giraffe and greater kudu) during different seasons (wet, early dry and late dry season). We collated the number of large herbivore sightings in nutrient enriched hotspots from camera trap photographs during the wet (January), early dry (June) and late dry (October) season. In addition, above ground grass biomass and height were measured and related to grazer sightings. Furthermore, we carried out an experiment on simulated grazing to determine the response of grass growth to repeated (three times during growth season) clipping. We tested the following hypotheses: 1) season affects use of nutrient enriched hotspots by large herbivores, 2) age of nutrient enriched hotspots affects its use by large herbivores, 3) aboveground grass biomass and height varies with time after kraal use 4) repeated grazing result in compensatory aboveground grass biomass. This study was conducted at Debshan, a mixed cattle-wildlife ranch, located in central Zimbabwe.
Materials & method

Study site

Debshan ranch is located in central Zimbabwe (29° 13'E, 19° 36'S; 1230m elevation) (Fig 1). It is a mixed cattle-wildlife ranch that covers an area of 800 km² and supports a diversity of large mammal species that include impala, Burchell’s zebra, warthog, African savanna elephant, northern giraffe and greater kudu. The study area is characterized by a catenal vegetation pattern, with most areas consisting of grassed bushland with patches of Miombo woodland [34]. The dominant woody species is Acacia karroo Hayne with the major grass species being Hyparrhenia filipendula (Hochst.) Stapf., Eragrostis curvula (Schrad.) Nees., Heteropogon contortus (L.) Roem. & Schult., Bothriochloa insculpta (Hochst. Ex A. Rich.), Digitaria milanjiana (Rendle) Stapf., and Panicum maximum Jacq. Mean annual rainfall is 612 mm, with a rainy season that runs from November to March and a dry season from April to October [34]. Mean annual temperature is 22.6˚C, with October (31.4˚C) the hottest month and July the coldest (8.5˚C).

Creation of nutrient enriched hotspots

The short duration overnight cattle kraaling system which created the nutrient enriched hotspots (alt. previously kraaled sites) was introduced to Debshan ranch in 2012. A herd of cattle (approximately 400) is kept overnight in a kraal (70m by 100m) set up in the rangelands for seven days before being moved to a new location. The kraal is made of heavy, steel posts and thick canvas sheeting and is 1.5m high [2]. Metal poles and wire cable are used to keep the boma canvas sheeting in place. Sites for placement of kraals are randomly selected and no trees are cut. The minimum distance between any two kraals is approximately two km. The cattle are guarded by herders who sleep in portable houses adjacent to each kraal overnight to protect them from predators. During daylight hours, herders direct cattle to grazing areas and drinking points close to the kraals. The kraals are used once for a period of seven days and have water supplied in a trough.
Camera traps setting

We deployed Cuddeback Attack/Attack IR digital scouting cameras \((n = 11)\), Cuddeback C (modular) and E model cameras \((n = 25)\) (Cuddeback Trail Camera company, India) infrared camera traps at eighteen locations in nutrient enriched hotspots (alt. previously kraaled sites) of varying ages (1, 2, 3 and 4 years) and control sites in surrounding vegetation at Debshan ranch between January and October 2017 (Fig 1). Three kraals were randomly selected for each year treatment with each having a control marked 300m away. Each kraal replicate and control site had two cameras. The cameras were mounted on tree trunks at one meter above the ground to detect medium- to large-bodied mammals [35]. Cameras were set for pictorial (single capture/minute) data capture for diurnal and nocturnal animal at a trigger speed interval of 60 seconds and each image displayed date (dd/mm/yy), time (hh:mm) and camera number(ID). Secure Digital (SD) memory cards and non-rechargeable batteries were replaced at two week interval. The pictorial data was downloaded from the SD cards and stored in folders labeled according to kraal age. Microsoft Excel version 2016 was used to store the photographic data with the following details: camera location (kraaled or unkraaled area), camera unit identifier, date (dd/mm/yy), time (hours, minutes) and animal species. Data collection was done in January \((n = 30\) days), June \((n = 30\) days) and October \((n = 30\) days) 2017. The number of animal sightings of each large herbivore species during each period was recorded from the camera trap data and expressed as number of animal sightings per day. All successive photographs of a species at the same camera were treated as independent if ten minutes passed with no captures of the particular species [36]. Camera trapping is non-intrusive and effective in studying large herbivore spatial and temporal use of habitats [37].

Aerial census data

Aerial censuses are conducted annually at Debshan ranch using a helicopter during July-August. Data for the year 2017 is presented in this study and used to calculate sighting indices for the six large herbivores studied.

Sighting indices

Sighting indices were calculated using the formula:

\[
\text{Sighting index} = \frac{\text{Number of wildlife sightings}}{\text{Total wildlife population}}
\]

Estimates of aboveground grass biomass cropped by grazers

We set up four chicken wire mesh (2 cm diameter holes) herbivore exclusion movable cages \((1m \times 1m \times 1m)\) in each nutrient enriched hotspot (alt. previously kraaled site) and surrounding vegetation to estimate aboveground grass biomass cropped by grazing herbivores. The cages were kept in the same position during the growth season (October 2016 to May 2017). The difference in aboveground grass biomass inside and outside the movable cages was assumed to represent grass cropped by the grazing herbivores [38]. Aboveground grass biomass both inside and outside the movable cages was clipped using a clipper, air dried, before oven drying at 60˚C for 48 hours and then weighed. All the aboveground grass inside the movable cage was clipped to ground level. Cropped aboveground grass biomass was then calculated as the difference between aboveground grass biomass inside and outside the mobile cages. Grass height was also measured in each sampling site using a tape measure to the nearest mm.

To test if repeated grazing results in compensatory aboveground grass biomass we clipped grass inside movable cages three times during the growth season and compared with aboveground grass biomass in a single clipping at the end of the growth season. Grass was clipped to
ground level on each occasion. Aboveground grass biomass inside movable cages located in nutrient enriched hotspots (alt. previously kraaled sites) of different ages (1, 2, 3 and 4 years after kraal use) were clipped three times at twenty-eight day intervals from the beginning of the growth season (28/01/2017) (1st clipping), peak of the growth season (25/02/2017) (2nd clipping) and at the end of the growth season (25/03/2017) (3rd clipping). The aboveground grass biomass removed at each of the three clipings was recorded and the sum for all the clipping calculated. Compensatory aboveground grass biomass ($gm^{-2}$) was calculated using the formula: total aboveground grass biomass for all three clipings–aboveground grass biomass clipped once at the end of growth season.

**Statistical analysis**

A total of 2833 camera images captured during the study period (90 days) were used for analysis of number of wildlife sightings [wet (January): n = 324; early dry (June): n = 874; late dry season (October): n = 1635]. The wildlife sightings were expressed as number of sightings per day to allow comparison of data for the three periods (wet, early dry and late dry season) as the number of camera trap images varied with season.

All data were tested for homogeneity of variance and normality using Levene statistics and Shapiro-Wilk test, respectively prior to statistical analyses. The effect of season (wet, early dry and late dry) and age of kraal (no kraaling, one, two, three and four years after kraaling) on number of wildlife sightings per day (proxy for use of nutrient enriched hotspots) were tested using one way analysis of variance. Above ground grass biomass, cropped grass biomass, clipped grass biomass and compensatory grass growth among the different aged kraal sites (no kraaling, one, two, three and four years after kraaling) were also compared using one way analysis of variance. Where differences among treatments were significant Tukey’s HSD was used for pairwise *post hoc* comparisons. Burchell’s zebra and warthog (grazers) number of sightings per day were related to aboveground grass biomass and height using Pearson square correlation. Sight- ing index was also related to wildlife population using Pearson square correlation. All data analysis was carried out using IBM SPSS 16.

**Results**

Zebra, warthog, impala, African savanna elephants, giraffe and greater kudu had large number of sightings per day in the camera traps to be used as representative species for the three feeding guilds (grazers, mixed feeders and browsers). Zebra and warthog are grazers, impala and elephants mixed feeders and giraffe and greater kudu obligate browsers. Other wildlife species with low numbers sighted in the camera traps were hare (*Lepus capensis*), common duiker (*Sylvicapra grimmia*), and steenbuck (*Raphicerus campestris*).

For each of the six species, there were no significant differences in sightings across the three seasons (early, early dry and late dry) (Burchell’s zebra: $F_{2,9} = 1.45, p = 0.285$; warthog: $F_{2,9} = 2.40, p = 0.146$; impala: $F_{2,9} = 0.72, p = 0.511$; African savanna elephant: $F_{2,9} = 0.02, p = 0.899$; giraffe: $F_{2,9} = 0.08, p = 0.926$; greater kudu: $F_{2,9} = 0.24, p = 0.795$) (*Fig 2*). Impala ($F_{4,10} = 11.06, p = 0.001$) and African savanna elephant ($F_{4,10} = 153.45, p < 0.001$) sightings per day significantly varied with time after kraal use (alt. age of nutrient enriched hotspots), while there were no significant differences for Burchell’s zebra ($F_{4,10} = 2.80, p = 0.09$), warthog ($F_{4,10} = 1.66, p = 0.24$), giraffe ($F_{4,10} = 2.11, p = 0.15$) and greater kudu ($F_{4,10} = 2.39, p = 0.12$) (*Fig 3*). Impala mostly used nutrient enriched hotspots one and four years after use, while African savanna elephant mostly preferred to use nutrient enriched hotspots one year after kraaling.
Aboveground grass biomass was highest in surrounding vegetation (unkraaled sites) \( (F_{4,20} = 1167, p < 0.001) \), with most aboveground grass biomass cropping occurring in the three year old nutrient enriched hotspots \( (F_{4,20} = 112.98, p < 0.001) \) (Fig 4). Grass was tallest in surrounding vegetation (unkraaled sites) \( (F_{4,20} = 407.13, p < 0.001; \) Fig 5). Zebra and warthog sightings were not significantly correlated to aboveground grass biomass (zebra: \( r = 0.54, p = 0.34, n = 5 \); warthog: \( r = -0.68, p = 0.21, n = 5 \)) and grass height (zebra: \( r = 0.57, p = 0.31, n = 5 \); warthog: \( r = -0.76, p = 0.14, n = 5 \)). Aerial census counts at Debshan ranch in 2017 showed that impala and giraffe were the most and least abundant wildlife species respectively (Fig 6). Giraffe had the highest sighting index \( (F_{5,18} = 7.02, p = 0.001; \) Fig 7). Sighting index was not significantly correlated to wildlife population \( (r = 0.10, p = 0.85, n = 6) \). Repeated grass clipping (proxy for grazing) resulted in compensatory above ground grass biomass, with the highest in the one year after kraal use sites (Table 1).

**Discussion**

Our research highlights the importance of using short duration overnight cattle kraaling in rangelands to create nutrient enriched hotspots attractive to wildlife in an African savanna ecosystem. We used the number of animal sightings per day from camera traps as a proxy for use of nutrient enriched hotspots and surrounding vegetation. All the six large herbivores (zebra, warthog, impala, African savanna elephants, giraffe and greater kudu) used nutrient enriched hotspots throughout the year. Previous studies have reported impala, warthog, African savanna elephants and other large herbivores as using nutrient enriched hotspots \[1,3,5,12\].
Our findings did not support the first hypothesis as wildlife use of nutrient enriched hotspots was similar during the wet and dry season. This suggests that the nutrient enriched hotspots provided nutritive forage and/or refuge to wildlife during both wet and dry season [1,39].

Our results showed that giraffe used nutrient enriched hotspots more than the surrounding vegetation. Conversely, Veblen and Porensky [5] reported giraffes as avoiding nutrient enriched hotspots, instead foraging in the surrounding vegetation. Our short duration overnight cattle kraaling system did not cut down trees and shrubs, with plant damage only due to cattle trampling. Traditional glades in east Africa are treeless because trees are cut down for use as kraal fences [40]. Zebra, African savanna elephants and kudu had low sighting indices, implying low use of nutrient enriched hotspots. Veblen and Porensky [5] also reported zebra as not actively seeking out high quality forage in nutrient enriched hotspots, presumably, because their large size and hind-gut fermentation allowed them to consume fibrous diets. In addition, low aboveground grass biomass in nutrient enriched hotspots, presumably, made them less attractive to zebra [4].

Our results showed that only mixed feeders use of nutrient enriched hotspot varied with time after kraal use (alt. age of previously kraaled sites). For both impala and African savanna elephant nutrient enriched hotspots were mostly used one year after kraal use. This was, presumably, due to the openness of these sites which improved impala predator detection and
Fig 4. Mean ($\pm$SE) aboveground and cropped grass biomass at nutrient enriched hotspots of different ages. Different letters (a, b, c and d—for aboveground grass biomass; and A, B, C, D—for cropped aboveground grass biomass) show differences in the treatments.

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Fig 5. Mean ($\pm$SE) grass height in nutrient enriched hotspots of different ages. Similar letters show that there were no significant differences between the treatments.

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allowed African savanna elephant easy movement. No plausible explanation could be proffered for the impala use of four year old nutrient enriched hotspots. Previous studies reported nutrient enriched hotspots as most attractive to mixed feeders [14,41,42]. Shannon et al. [42] attributed this to the ability of mixed feeders (particularly African savanna elephants) to mainly browse while also consuming grass. Huruba et al. [1] reported cattle as breaking woody plant stems and stripping them of foliage during overnight kraaling initiating resprouting, with the resprouts attractive to impala due to high foliar nitrogen and low condensed tannin concentrations.

Fig 6. Number of wildlife at Debshan ranch in 2017.

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Fig 7. Mean (±SE) sighting index for six wildlife species. Similar letters show that there were no significant differences between the treatments.

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Table 1. Mean (±SE) aboveground grass biomass (g m⁻²) in cages clipped once and repeatedly (three times) in previously kraaled sites of different ages.

|                  | clipped once | Repeated clipping | 2nd clipping | 3rd clipping | Total for repeated clipping | Compensatory growth |
|------------------|--------------|-------------------|--------------|--------------|-----------------------------|---------------------|
|                  |              |                   |              |              |                             |                     |
| 1 year after kraal use | 248.40 ± 4.46 | 351.20 ± 2.58 | 355.20 ± 1.83 | 353.00 ± 4.36 | F₁₆ = 1964.00, p < 0.001    | 351.20 ± 3.31       |
| 2 years after kraal use | 252.40 ± 3.75 | 379.20 ± 2.85 | 72.60 ± 0.93  | 497.40 ± 1.81 | F₁₆ = 101.90 ± 1.58         | 274.20 ± 1.66       |
| 3 years after kraal use | 287.40 ± 3.33 | 331.20 ± 15.99 | 71.05 ± 2.52  | 353.00 ± 4.36 | F₁₆ = 3.33 ± 0.95           | 209.40 ± 2.58       |
| 4 years after kraal use | 255.20 ± 4.19 | 281.60 ± 2.20 | 37.00 ± 2.10  | 353.00 ± 4.36 | F₁₆ = 155.50 ± 1.62         | 101.90 ± 1.58       |

Aboveground grass biomass and height were highest in surrounding vegetation, presumably, because of low grazing intensity. This is supported by the low aboveground grass biomass cropping in the surrounding vegetation. These results show that nutrient enriched hotspots were more intensely grazed than surrounding vegetation. Huruba et al. [2] reported warthogs as intensely grazing in nutrient enriched hotspots. Interestingly, aboveground grass biomass was not significantly correlated to zebra and warthog number of sightings per day. Generally, aboveground grass biomass and height in nutrient enriched hotspots tended to decrease with time after kraal use (see Fig 4), with cropping showing an opposite trend. Aboveground grass biomass cropped was within the range of 89 to 951 g m⁻² reported in the Kruger National Park by Burkepile et al. [43]. Grass in the surrounding vegetation was moribund and unattractive even to zebra that are tolerant to fibrous diets. Zebra are tolerant to fibrous low quality grass because of their fast passage rate of forage through the gastrointestinal tract [44].

Repeated clipping (proxy for grazing) resulted in compensatory aboveground grass biomass in nutrient enriched hotspots. McNaughton [27] reported grasses in the Serengeti as overcompensating lost foliage. Compensatory aboveground grass biomass declined with time after kraal use, presumably, due to a decline in soil fertility. Improved soil fertility due to dung and urine deposition, particularly one year after kraal use, could have enhanced grass compensatory growth. Venter et al. [45] reported nutrient addition (in the form of animal dung) as increasing aboveground grass biomass. The decline in aboveground grass biomass regrowth between first and third clipping (see Table 1) was, presumably, due to resource exhaustion as a result of multiple grass resprouting in response to clipping [46]. Mudongo et al. [47] also reported a decrease in aboveground grass biomass regrowth with increasing clipping frequency. In the long-term repeated grazing could negatively affect tillering leading to the loss of the grass [48]. While previous studies have shown that grazing in the preceding growth season reduces grass productivity in the next growth season [49,50], our results show that a decline in regrowth occurs in the current season. In addition, the decline in aboveground grass biomass regrowth with repeated clipping could be attributed to reduced soil moisture availability with advancing growth season which negatively affects nutrient mineralization [51]. Soil mineralization is higher early in the growth season when soil moisture is high as compared to late in the dry season [47].

The purpose of calculating sighting index was to determine if use of nutrient enriched hotspots was influenced by wildlife abundance. Thus, the fact that sighting index was not significantly correlated to wildlife population implies that use of nutrient enriched hotspots was independent of animal abundance. For example, giraffe had the lowest population and highest sighting index, implying that there were frequent users of nutrient enriched hotspots.

**Conclusion**

Our findings showed that nutrient enriched hotspots created through short duration overnight cattle kraaling in natural rangelands were attractive to large herbivores. Large herbivore use of
nutrient enriched hotspots was similar during wet and dry season. Time after kraal use (alt. age of previously kraaled sites) had an effect on mixed feeders (impala and African savanna elephants) use of nutrient enriched hotspots but not grazers (zebra and warthog) and browsers (giraffe and greater kudu). Aboveground grass biomass and height in nutrient enriched hotspots was lower than in surrounding vegetation due to more intense grazing by large herbivores. Repeated grazing resulted in compensatory grass growth that declined with age of nutrient enriched hotspots. The creation of nutrient enriched hotspots in rangelands improves the availability of foraging resources to large herbivores in both wet and dry season.

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References

1. Huruba R, Mundy PJ, Sebata A, Purchase GK, MacFadyen DN. Impala, Aepyceros melampus: does browse quality influence their use of sites originally utilized as short-duration kraals in a southern African savanna? Rangeland J. 2017; 39: 113–121.
2. Huruba R, Mlambo T, Mundy P.J., Sebata A, MacFadyen D.N., 2018. Short duration overnight cattle kraaling in natural rangelands: Implications for grass composition, quality, above ground biomass, species diversity and basal cover. Agric., Ecosyst. Environ. 2018; 257: 144–151.
3. Sibanda P, Sebata A, Mufandaedza E, Mawanza M. Effect of short-duration overnight cattle kraaling on grass production in a southern African savanna. Afr. J. Range For. Sci.2016 ; 33: 217–223.
4. Porensky LM, Veblen KE. Generation of ecosystem hotspots using short-term cattle corrals in an African Savanna. Range. Ecol. Manage. 2015; 68: 131–141.
5. Veblen KE, Porensky LM. Thresholds are in the eye of the beholder: plants and wildlife respond differently to short-term cattle corrals. Ecol. Appl. 2019; 00(00):e01982. https://doi.org/10.1002/eap.1982 PMID: 31348560
6. Ford AT, Goheen JR, Ottieno TO, Bidner L, Isbell LA, Palmer TM, et al. Large carnivores make savanna tree communities less thorny. Science. 2014; 346: 346–349. https://doi.org/10.1126/science.1252753 PMID: 25324387
7. Scholes R.J., Archer S.R., 1997. Tree-grass interactions in savannas. Ann. Rev. Ecol. Syst. 1997; 28: 517–544.
8. Augustine DJ, Frank DA. Effects of migratory grazers on spatial heterogeneity of soil nitrogen properties in a grassland ecosystem. Ecology. 2001; 82: 3149–3162.

9. Wolf A, Doughty CE, Malhi Y. Lateral diffusion of nutrients by mammalian herbivores in terrestrial ecosystems. PLoS ONE. 2013; 8: 1–10. https://doi.org/10.1371/journal.pone.0071352 PMID: 23951141

10. Augustine DJ, McNaughton SJ, Frank DA. Feedbacks between soil nutrients and large herbivores in a managed savanna ecosystem. Ecol. Appl. 2003; 13: 1325–1337.

11. Russ W, McNaughton SJ. Grazing and the dynamics of nutrient and energy regulated microbial processes in the Serengeti Tanzania Kenya grasslands. Oikos. 1987; 49: 101–110.

12. Augustine DJ, Veblen KE, Goheen JR, Riginos C, Young TP. Pathways for positive cattle–wildlife interactions in semi-arid rangelands. Smithsonian Contributions to Zoology. 2011; 55–71. https://doi.org/10.5479/si.00810282.632.55

13. Donihue CM, Porensky LM, Foufopoulos J, Riginos C, Pringle RM. Glade cascades: indirect legacy effects of pastoralism enhance the abundance and spatial structuring of arboreal fauna. Ecology 2013; 94(4), 827–837.

14. Veblen K.E., 2012. Savanna glade hotspots: plant community development and synergy with large herbivores. J. Arid Environ. 2012; 78: 119–127.

15. Owen-Smith N. Novellie P. What Should a Clever Ungulate Eat? American Naturalist 1982; 119(2): 151–178. https://doi.org/10.1086/283902/

16. Codron D, Codron J. Reliability of delta C-13 and delta N-15 in faeces for reconstructing savanna herbivore diet. Mammalian Biol. 2009; 74: 36–48.

17. Codron J, Codron D, Lee-Thorp JA, Sponeheimer M, Kirkman K, Duffy KJ, et al. Landscape-scale feeding patterns of African elephant inferred from carbon isotope analysis of faeces. Oecologia. 2011; 165: 89–99. https://doi.org/10.1007/s00442-010-1835-6 PMID: 21072541

18. Mayengo G, Piel AK, Treydte AC. The importance of nutrient hotspots for grazing ungulates in a Miombo ecosystem, Tanzania. PLoS ONE. 2020; 15(3)2020 : e0230192. https://doi.org/10.1371/journal.pone.0230192 PMID: 32226036

19. Augustine DJ. Long-Term, Livestock-Mediated Redistribution of Nitrogen and Phosphorous in an East African Savanna. Journal of Applied Ecology 2003; 40:137–149.

20. Treydte AC, Baumgartner S, Heitkönig IM, Grant CC, Getz WM. Herbaceous forage and selection patterns by ungulates across varying herbivore assemblages in a South African savannah. PLoS One. 2013; 8: e82831. https://doi.org/10.1371/journal.pone.0082831 PMID: 24358228

21. Muchiru AN, Western DJ, Reid RS. The role of abandoned pastoral settlements in the dynamics of African large herbivore communities. J. Arid Environ. 2008; 72: 940–952.

22. Young C, Fritz H, Smithwick EAH, Venter JA. Patch-scale selection patterns of grazing herbivores in the central basalt plains of Kruger National Park. Afr. J. Range For. Sci. 2020; 1–15.

23. Claus M, Frey R, Kiefer B, Lechner-Doll M, Loehlein W, Polster C, et al. The maximum attainable body size of herbivorous mammals: morphophysiological constraints on foregut and adaptations of hindgut fermenters. Oecologia. 2003; 136: 14–27. https://doi.org/10.1007/s00442-003-1254-z PMID: 12712314

24. Fleurance G, Duncan P, Fritz H, Gordon IJ, Grenier-Loustalot MF. Influence of sward structure on daily intake and foraging behaviour by horses. Animal: An Int. J. Anim. Biosci. 2010; 4(3); 480–485.

25. Arsenault R, Owen-Smith N. Resource partitioning by grass height among grazing ungulates does not follow body size relation. Oikos. 2008; 117: 1711–1717.

26. Laca EA, Ungar ED, Seligman N, Demment MW. Effects of sward height and bulk density on bite dimensions of cattle grazing homogeneous swards. Grass Forage Science. 1992; 47: 91–102.

27. McNaughton S. Grazing as an optimization process: grass-ungulate relationships in the Serengeti. Am. Nat. 1979; 113: 691–703.

28. McNaughton S. Compensatory plant growth as a response to herbivory. Oikos. 1983; 40; 329–336.

29. Fuls ER. Semi-arid and arid rangelands: A resource under siege due to patch-selective grazing. J. Arid Environ. 1992; 22 (2): 191–193.

30. Hilbert DW, Swift DM, colouring JK, Dyer MI. Relative growth rates and the grazing optimization hypothesis. Oecologia. 1991; 85: 14–16. https://doi.org/10.1007/BF00344645 PMID: 28310002

31. Belsky AJ. Does herbivory benefit plants? A review of the evidence. Am. Nat. 1986; 127: 870–892.

32. Hawkes CV, Sullivan JJ. The impact of herbivory on plants in different resource conditions: a meta-analysis. Ecology. 2001; 82: 2045–2058.

33. Hobbs N.T., 1996. Modification of ecosystems by ungulates. J.Wildl. Manage. 1996; 60: 695–713.
34. Dunham KM, Robertson EF, Swanepoel CM. Population decline of tsessebe antelope (*Damaliscus lunatus lunatus*) on a mixed cattle and wildlife ranch in Zimbabwe. Bio. Cons. 2003; 113: 111–124.

35. Meek DP, Ballard AG, Falzon G. The higher you go the less you will know: placing camera traps high to avoid theft will affect detection. Remote Sensing in Ecology and Conservation. 2016; https://doi.org/10.1002/rse2.32 PMID: 31423326

36. Kolowski JM, Forrester TD (2017) Camera trap placement and the potential for bias due to trails and other features. PLoS One. 2017; 12(10): e0186679. https://doi.org/10.1371/journal.pone.0186679 PMID: 29045478

37. Carbone C, Christie S, Conforti K, Coulson T. The use of photographic rates to estimate densities of tigers and other cryptic mammals. Anim. Conserv. For. 2001; 4(1): 75–79.

38. McNaughton SJ, Milchunas DG, Frank DA. How can net primary productivity be measured in grazing ecosystems? Ecology. 1996; 77: 974–977.

39. Anderson TM, Hopcraft JGC, Eby S, Ritchie M, Grace JB, Off H. Landscape-scale analyses suggest both nutrient and antipredator advantages to Serengeti herbivore hotspots. Ecology. 2010; 91: 1519–1529. https://doi.org/10.1890/09-0739.1 PMID: 20503883

40. Porensky LM, Veblen KE. Grasses and browsers reinforce landscape heterogeneity by excluding trees from ecosystem hotspots. Oecologia. 2012; 168: 749–759. https://doi.org/10.1007/s00442-011-2123-9 PMID: 21935665

41. van der Waal C, Kool A, Meijer SS, Kohi E, Heitkönig IM, de Boer WF, et al. Large herbivores may alter vegetation structure of semi-arid savannas through soil nutrient mediation. Oecologia. 2011; 165: 1095–1107. https://doi.org/10.1007/s00442-010-1899-3 PMID: 21225433

42. Shannon G, Mackey RL, Slatow R. Diet selection and seasonal dietary switch of a large sexually dimorphic herbivore. Acta Oecologica.2013; 46: 48–55.

43. Burkceple DE, Fynn RWS, Thompson DI, Lemoine NP, Koemer SE, Eby S, et al. Herbivore size matters for productivity–richness relationships in African savannas. J. Ecol. 2017; 105: 674–686.

44. Demment MW, van Soest PJ. A nutritional explanation for body-size patterns of ruminant and non-ruminant herbivores. Am. Nat. 1985; 125: 641–672.

45. Venter ZS, Hawkins H-J, Cramer MD. Does defoliation frequency and severity influence plant productivity? The role of grazing management and soil nutrients. Afr. J. Range For. Sci. 2020; https://doi.org/10.2989/10220119.2020.1766565

46. Werner PA, Prior LD. Demography and growth of subadult savanna trees: interactions of life history, size, fire season, and grassy understory. Ecol. Monographs. 2013; 83: 67–93.

47. Mudongo EI, Fynn RWS, Bonyongo MC. Role of herbivore impact and subsequent timing and extent of recovery periods in rangelands. Range. Ecol. Manage. 2016; 69: 327–333.

48. Tomlinson KW, O’Connor TG. Control of tiller recruitment in bunchgrasses: unifying physiology and ecology. Funct. Ecol. 2004; 18: 489–496.

49. Kirkman K. The influence of various types and frequencies of rest on the production and condition of sour veld grazed by sheep or cattle. Afr. J. Range Forage Sci. 2002; 19: 93–105.

50. Turner CL, Seaestadt TR, Dyer MI. Maximization of aboveground production in grasslands: the role of defoliation frequency, intensity and history. Ecol. Appl. 1993; 3: 175–186. https://doi.org/10.2307/1941800 PMID: 27759215

51. Fierer N, Schimel JP. Effects of drying–rewetting frequency on soil carbon and nitrogen transformations. Soil Biol. Biochem. 2002; 34: 777–787.