Spatial–temporal changes in the degradation of marshes over the past 67 years
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Agricultural reclamation is widely regarded as a primary cause of marshes degradation. However, the process of marshes degradation on different geomorphology has rarely explored, which fail to explain the marshes degradation driven by natural restrictions in detail. The information deficiency unable propounded the targeted suggestions for the sustainable management of marshes. According to the development of China, we quantified the degradation rate of marshes on different geomorphic types from 1954 to 2020 in a typical transect in the Sanjiang Plain. The results indicated that (1) A total of 1633.92 km² of marshes reduced from 1954 to 2020. And 97% (1582.35 km²) of marshes were converted to crop cultivation. The process of marshes degradation had obvious historical stages characteristics. The marshes degradation rate showed a trend of increasing first and then decreasing. The most serious period was 1995–2005 (6.29%) which was approximately 35 times of the period of before the reform and opening up (1954–1976) a minimal shrunk period. (2) The background of geological tectonic decided the whole trends in marshes degradation process. The degradation occurred first and worst in the meco-scale recent slow ascent region, and then extended to substantially recent slow subsidence region and the small-amplitude recent slow ascent region. (3) Significant location characteristics of marshes degradation reflected in this research. The spatial location of marshes degradation on the sub-regions sequentially consisted of alluvial plain, lower terrace, high floodplain, micro-knoll, low floodplain, and depressions. (4) Most of the existing marshes of the sub-Sanjiang Plain distribution in the national reserves. This study provides important scientific information for restoration and conservation of marshes.

As the most productive and economically valuable ecosystems1,2, Marshes offer multiple ecosystem services, including climate regulation, flood control, carbon storage and so on3–5. Simultaneously, they play an important role in biodiversity conservation and regional ecological balance, etc6–7. In fact, wetlands contribute directly or indirectly to 75 Sustainable Development Goal indicators of the United Nations Environment Programme (UNEP)8. However, with the developing society and the increasing population, wetlands are also suffering unprecedented destruction. For example, China9–11, Italy12, Brazil13, and global monitoring of wetland14. Therefore, many researchers in different spatial–temporal scale and wetlands types about spatial–temporal characteristic of wetlands change was implement15–17. Long-term wetland monitoring studies have shown that the average annual loss rate of wetlands is as high as 50% and human activities are the main cause of wetland degradation, in which agricultural activities are the leading anthropogenic force driving factors5,13,14.

However, the degradation process of wetlands is not only driven by human activities but also restricted by the natural environment18,19. Studies have shown that geomorphology is an important factor that affects the distribution and genesis of wetland16–22. For example, over-humidity or waterlogging on the surface is the key feature to the formation of marshes, the non-zonal nature of which indicates that the hydrological characteristics of a region are limited by geomorphic conditions23. Mitchell. M24 also stated the importance of the marsh loss to overall estuarine function may depend on the location and type of marsh lost. With the sea-level rise, knowing which marshes are most vulnerable allows for the prioritization of restoration and conservation efforts, minimizing future impacts to estuarine systems. Therefore, understanding the evolution process of wetlands driven by geomorphic restrictions and human activities is of great significance for formulating long-term wetland management strategies and revealing the temporal and spatial characteristics of wetland evolution16–25.26.

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China's national conditions determine that China's wetlands degradation has its own national historical characteristic, simultaneously, follows the constraints of nature in space. The Sanjiang Plain is the largest marsh area in China, formed by the alluviation of the Heilongjiang, Nongjiang, and Wusuli Rivers. Since 1950s, due to the flat landform and fertile land, now, it has become China's main commodity grain base with the support of the national reclamation policies. However, the serious shrinkage of marshes in the Sanjiang Plain has caused most of the ecological functions lost. Although several ecological conservation policies and management measures were formulated to protect and restore natural ecosystems in this region, the marshes in the Sanjiang Plain has still been reclaimed for farmland constantly. So, a comprehensive study focusing on the response of marshes degradation to policies and natural constraints is urgently needed to improve the efficiency of marshes restoration and protection.

The primary objective of this study is to explore the spatial–temporal characteristic of marshes degradation in sub-Sangjiang Plain with geomorphic restrictions and human driven from 1954 to 2020 combine with RS and GIS technology. Specifically, we aimed to (1) exam the landscape pattern changes of sub-Sangjiang Plain in 1954–2020, (2) quantify the degradation of marshes in sub-Sangjiang Plain in different China development historical stage, and (3) explore the degradation of marshes in geographic types. We expect to provide scientific basis and theoretical support for the conservation, restoration and sustainable development of marshes in China.

**Study area**

The study was conducted in a subset of Sanjiang Plain, called sub-Sanjiang Plain. This region is the center area of Sanjiang Plain in Heilongjiang Province, northeast China, which locates north of Wanda Mountain (Fig. 1). The Sanjiang Plain is a low alluvial plain formed by three rivers: Heilongjiang River, Songhua River, and Wusuli.
River. The total area is 16,073.7 km². The climate is temperate continental with four distinct seasons, including 6 months of freezing conditions. The mean annual precipitation is 603.8 mm and mainly concentrates from June to August. The long freezing periods and clayey soil prevent water from infiltrating to deep layer. The geological structure type is Tongjiang Depression. The low-lying geomorphologic pattern in the area has been shaped by the surface subsidence since the Quaternary. The geomorphology types mainly included riparian zone and terrace. Besides, a small amount of residual hill scatter in the vast low plain and the micro-landform are well developed on the riparian zones. The densely covered river, low and even terrain, and poor drainage rise the water table level in flood season. All these natural and geographical conditions promote the formation and development of the marshes. Furthermore, the study area is suitable for the growth of annual crops such as corn, soybean and rice, etc. And the main soil types are meadow soil, swamp soil and albic soils.

Data preparation and methods

Many previous studies have shown that wetlands change in China have obvious historical characteristics. Reform and opening up is an important turning point of China’s development. China’s policy formulation, economic, agricultural and environmental development and changes were closely related to the basic national policy of reform and opening up. Therefore, this study is based on the large social background of reform and opening up. Considering the availability of data and the lag of policy implementation, the land-use/cover datasets (1:100,000) of the typical transect zone in 1954–2020 were chose. The datasets derived from Northeast Institute of Geography and Agroecology (NEIGAE), Chinese Academy of Sciences (CAS) (http://www.iga.ac.cn) and USGS Global Visualization Viewer (https://glovis.usgs.gov/) (Table 1). More detailed time points and reasons were shown in Table 1. In addition, an auxiliary data of 1:500,000 geomorphic map compiled by the Institute of Geographic Sciences and Natural Resources Research, Chinese Academy of Sciences (https://www.resdc.cn/data.aspx?DATAID=260).

Data preparation. To meet the purpose of this study, according to the land-use/cover classification system (Table 2) we adjusted the existing achievement land-use/cover data of the NEIGAE, CAS: 1954, 1976, 1985, 1995, 2000 and 2005, firstly. After preprocessing the remote sensing images in 1982, 1990, 2010, 2015 and 2020.
Then the corresponding land-use data were obtained with visual interpretation using ArcGIS. The land-use/cover data of 1982 was generated by updating the adjusted land-use/cover maps of 1976 and 1985 with Landsat Multispectral Scanner (MSS) image in 1982. The land use data of 1990 was visual interpreted by the Landsat Thematic Mapper (TM) image with the updating to the adjusted land-use/cover maps of 1985 and 1995. The land use data in 2010, 2015 and 2020 were derived from updating adjusted the land-use/cover map of 2005 with the Landsat Thematic Mapper (TM) image and Landsat 8 OLI (Operational Land Imager) and TIRS (Thermal Infrared Sensor) image, respectively. In addition, the overall accuracy of land-use was reached 90.5%.

Methods. The criterions of transect establishment. The transect is an important method to study the spatial differences, which has been widely used in the fields of geography and ecology. The based-transect method has become an important research tool to explore the zonal variation law in International Geosphere-Biosphere Program (IGBP). In order to accurately explore the temporal-spatial process and characteristics of the marsh loss and degradation in the sub-Sanjiang Plain, the criterions of the establishment of the typical transect are as follows:

1) The transect is essential to cover all geomorphic types of the distribution area of the marshes in the sub-Sanjiang Plain;
2) The transect is essential to cover the main factors of the formation of the marshes;
3) The transect is able to describe the characteristics of the marsh changes in the sub-Sanjiang Plain from 1954 to 2020;
4) The transect region is able to represent the land use patterns of sub-Sanjiang Plain.

The condition of typical transect. In this study, we selected a typical transect with the size of 1905 km² from the sub-Sanjiang Plain (Fig. 1). Four main rivers flow through the transect including Heilongjiang River, Nongjiang River, Bielahong River and Wusuli River. The established transect mainly depends on the following factors:

1) The geomorphic types of the transect include alluvial depression, middle bar, alluvial plain, micro-knoll, low floodplain, high floodplain and lower terrace etc. Figure 2, which cover the main geomorphic types of marsh distribution in the sub-Sanjiang Plain;
2) The transect zone is shaped by the Tongjiang Depression. The reasons for the formation of the marsh in the sub-Sanjiang Plain are covered in this transect zone, including low-lying terrain, poor surface drainage and soil water penetration, well developed micro-geomorph, humid climate and the dense plain river;
3) From 1954 to 2020, the marsh area has decreased by 82% from 1806.7 to 245.7 km², which indicates that there is a severe degradation of marshes;
4) The landscape pattern was changed from the dominated marsh to the dominated farmland during the study period, which is able to represent the changes of land use pattern in the sub-Sanjiang Plain.
5) There are two main nature reserves in the transect zone, which are Sanjiang Heilongjiang National Nature Reserve and Bacha Island Heilongjiang National Nature Reserve (Fig. 2).

Segmentation of transect zone. Although the general trend of the neotectonics on sub-Sanjiang plain was descending, it was accompanied by intermittent rising in the process of subsidence, which was mainly controlled by the north-east secondary fault. And due to the surface subsidence since the Quaternary, the geological structure type of the typical transect zone is Tongjiang Depression. The specific manifestations are the Heilongjiang_Nongjiang substantially recent slow subsidence region, the Nongjiang_Bielahong River meco-scale recent slow ascent region and the Bielahong River_Wusuli River small-amplitude recent slow ascent region. Combination of the geological formation movement, the typical transect was divided into three regions named first-region, including Heilongjiang_Nongjiang Region (I), Nongjiang_Bielahong River Region (II) and Bielahong River_Wusuli River Region (III) to reveal the spatial characteristics of marsh degradation. Then each region was further segmented into some sub-regions with the reference to fluvial geomorphic types, such as river region, middle bar region, depression region, alluvial plain region, micro-knoll region, lower terrace region, low floodplain region, high floodplain region and low hill region (Fig. 2).

Quantification of wetland degradation. For well understanding the rate of regional marshes changes and their characteristics differences, the marshes dynamic degree was proposed. Wetland dynamic degree can be determined by the land use dynamic degree.

The equation is as follows:

\[ LC = \frac{U_b - U_a}{U_a \cdot T} \times 100\% \]  

where \( U \) is the marsh area (km²), \( b \) represents the current year and \( a \) represents the past years, \( LC \) is the dynamic degree (annual shrinkage ratio) of marshes area in \( T \) years.

Statistical analysis. In this study we mapped the spatial pattern in typical transect and calculated the conversion among farmland and marshes by ArcGIS software. For a better understanding of the patterns and process in marshes degradation, we quantified both the area and annual shrinkage ratio of marshes (LC) in six stage (Table 3). Specifically, we gathered statistic and compared the spatial–temporal variable changes among the first- regions and sub-regions.
Results

Spatial–temporal characteristics of marsh degradation. As Fig. 3 shown that the landscape patterns in the typical transect zone have dramatically changed during the 67 years, the landscape matrix from marshes changed into farmland. As Fig. 4 shown that paddy appeared around 2000 and gradually increased eventually, became the main landscape type in sub-Sangjiang Plain. Before this, marshes were only reclaimed into dry farmland. A total of 1633.92 km² of marshes reduced from 1954 (1806.71 km²) to 2020 (172.9 km²). Meanwhile, there was an increase of 1582.35 km² (82%) of crop cultivation, of which dry farmland and paddy fields were 332 km² (17%) and 1250.35 km² (65%) respectively. Additionally, the loss of marshes caused by agricultural encroachment were different in the first-region. Which firstly occurred in I region and then expanded to other two regions (Fig. 5). Compared to region I and III, the marsh degradation rate in region II was faster and worse. In the start of 1954, the areas of marshes in the first-region were 599.18 km², 604.78 km² and 602.70 km² respectively (Fig. 5) but were 100.06 km², 8.36 km² and 64.32 km² in 2020.

Table 3. Description of historical stage.
Quantification of the marsh degradation was shown in Table 4. The annual shrinkage ratio was continuously increased throughout the whole period which was rose rapidly at a rate of more than 3 times of each history stage in 1954–2005, but started slowly in the period of deepening reform and opening up until 2020. The marsh degradation rate was more notable in the period of transformation reform and opening up stage (1995–2005), which was approximately 35 times of the period of before the reform and opening up (1954–1976) a minimal shrunk period.

**Dynamic change degree of marsh.** The annual degradation rate of marshes in regions I, II and III (Table 5) was all fluctuation, however, the trends of each region has its own characteristic. The annual degradation ration of marshes in regions I and II were approximately 4 times and 3 times of the rate respectively.
from 1954 to 2005, and both were reached the highest value in 1995–2005. However, the rate of region II in each observation stage was both higher than regions I and III. In terms of the highest value, the max annual degradation rate in region II was 8.56%, which was 1.3 and 1.5 times in regions I and III, respectively, although the annual degradation rate decreased in the subsequent period, it was always greater than the highest values of regions I and III; then the annual degradation rate in region I was fall to 1.77% in 2005–2015 which was the minimum of the contemporaneity, and rose to almost the highest rate of region I in the next 5 years. Simultaneously, the annual degradation rate of marsh in region III declined slightly after continue growing from 1954 to 2015. In fact, after the period of the deepening of reform and opening up, the average annual degradation rate has remained high although it has changed.

**Spatial location characteristics of marsh change.** *Heilongjiang_Nongjiang region.* As with Fig. 6 and Table 6 shown, the marshes in District Heilongjiang_Nongjiang region were mainly distributed on high floodplains (202.7333 km²), low floodplains (262.992 km²), lower terrace (123.456 km²), and depression (7.392 km²). The sporadic degradation of marshes first occurred on the high floodplain (1954–1976), and then appeared in the lower terrace (1976–1982). When the agriculture reclamation was extended to the floodplain (1990–1995), the marshes on depressions were also began to be continuously cultivated. However, after only 10 years, the marshes on depressions disappeared. It was worth mentioning that before 2005, the mashes were mainly converted into dry land but paddy fields after 2005. The intrusion of agriculture on various geographic was differed

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**Table 4.** Annual degeneration rate of the marsh in typical transect.

| Stage name                                      | Time intervals | Annual shrinkage ratio (LC) (%) |
|-------------------------------------------------|----------------|---------------------------------|
| The period of before the reform and opening up  | 1954–1976      | 0.2                             |
| Initial period of reform and opening up         | 1976–1985      | 0.67                            |
| The period of comprehensive reform and opening up| 1985–1995      | 1.89                            |
| The period of transformation reform and opening up| 1995–2005      | 6.29                            |
| The period of deepening reform and opening up   | 2005–2015      | 4.96                            |
| The period of deepening reform and opening up   | 2015–2020      | 5.94                            |

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**Table 5.** The annual variance ratio of the marsh degradation based on the first level division.

| Stage name                                      | Time intervals | Annual shrinkage ratio (LC) (%) | I      | II     | III    |
|-------------------------------------------------|----------------|--------------------------------|--------|--------|--------|
| The period of before the reform and opening up  | 1954–1976      | 0.08                           | 0.39   | 0.13   |
| Initial period of reform and opening up         | 1976–1985      | 0.38                           | 1.51   | 0.18   |
| The period of comprehensive reform and opening up| 1985–1995      | 1.41                           | 3.07   | 1.38   |
| The period of transformation reform and opening up| 1995–2005      | 6.39                           | 8.56   | 4.69   |
| The period of deepening reform and opening up   | 2005–2015      | 1.77                           | 7.21   | 6.7    |
| The period of deepening reform and opening up   | 2015–2020      | 6.22                           | 7.43   | 5.24   |
in every stage. The periods of most degradation of marshes on the high floodplain from 1985–2005 were reduced by 63%; lower terraces were 90% in 1990–2005; and low floodplains were 77% in 1995–2020. By 2020, the marshes distributed on the lower terraces was less than 0.03 of the original; the high and low floodplains were only 2%. Nongjiang_Bielahong region.

As Fig. 7 and Table 6 shown that marshes were mainly distributed on alluvial plains (299.916 km²), lower terrace (256.839 km²), followed by depressions (41.99 km²), and only a few were distributed on low floodplains (5.65 km²) and few micro-knoll (0.38 km²). The reclamation of agriculture on the low floodplain was mainly happened 1982–2000 and the marshes on it were reduced by 89%; alluvial plain were 85% in 1954–2000; lower terraces were 89% in 1976–2005; and depression were 89% in 1990–2005. By 2020, the marshes distributed on the Nongjiang-Bielahong river region were less than 10 km², almostly disappeared. The degradation process of marshes was different among different geographic. The Fig. 7 showed that the reclamation of farmland first started with large-scale reclamation on alluvial plain and lower terrace (1954–1976), then sporadic reclamation on low floodplain (1976–1982), followed by depressions, finally micro-knoll (1985–1990). With the minimum area of marshes (0.378 km²), micro-knoll occupied by farmland later but faster than other geographic and completely converted into farmland during 2000–2005. Compared with other geometric types, the process of marshes on depression eroded by farmland is the slowest, which was completely occupied by farmland until 2020.

Table 6. The distribution of marsh during 1954–2020 based on geomorphic types in the first level division.

| First-region | Sub-region                     | Percentage of marshes area in 1954 (%) |
|--------------|--------------------------------|----------------------------------------|
| I Low floodplain | 1 | 1.00 1.00 1.00 1.00 0.97 0.78 0.47 0.42 0.34 0.20 |
| I High floodplain | 1 | 0.96 0.97 0.90 0.76 0.75 0.47 0.27 0.26 0.27 0.20 |
| I Depression | 1 | 1.00 1.00 1.00 1.00 0.99 0.44 0.00 0.00 0.00 0.00 |
| I Lower terrace | 1 | 1.00 0.99 0.98 0.93 0.69 0.18 0.05 0.05 0.02 0.03 |
| I Middle bar | 0 | 0.00 0.00 0.00 0.00 0.00 0.00 0.00 0.00 0.00 0.00 |
| II Low floodplain | 1 | 0.86 0.91 0.79 0.56 0.49 0.02 0.02 0.01 0.01 0.00 |
| II Alluvial plain | 1 | 0.87 0.71 0.71 0.62 0.55 0.15 0.08 0.03 0.02 0.01 |
| II Depression | 1 | 1.00 0.95 0.95 0.91 0.83 0.53 0.20 0.07 0.02 0.00 |
| II Micro-knoll | 1 | 1.00 1.00 1.00 0.69 0.69 0.23 0.00 0.00 0.00 0.00 |
| II Lower terrace | 1 | 0.95 0.89 0.86 0.60 0.50 0.16 0.06 0.04 0.03 0.02 |
| II Low floodplain | 1 | 1.00 1.00 0.99 0.97 0.94 0.92 0.89 0.61 0.48 0.48 |
| II Alluvial plain | 1 | 0.96 0.96 0.95 0.88 0.79 0.53 0.32 0.13 0.03 0.00 |
| II Depression | 1 | 1.00 1.00 1.00 0.99 0.98 0.89 0.53 0.24 0.08 0.04 |
| II Micro-knoll | 1 | 1.00 1.00 0.99 0.79 0.52 0.22 0.02 0.01 0.01 0.00 |
| II Lower terrace | 1 | 0.93 0.93 0.89 0.81 0.69 0.46 0.31 0.15 0.07 0.03 |
| II Middle bar | 1 | 1.00 1.04 1.00 0.00 0.00 0.00 0.00 0.00 0.00 0.47 |
| II Low hill | 1 | 0.63 0.63 0.54 0.54 0.54 0.52 0.34 0.00 0.00 0.00 |
Bielahong_Wusuli region. The main geomorphic types of Bielahong River_Wusuli river region were alluvial plain, depressions, lower terrace, low floodplains, high floodplains, and micro-knoll. Sporadic farmland reclamation first occurred in alluvial plain, lower terrace, and high floodplain (1954–1976). After 1982, the sporadic reclamation began on micro-knoll and low floodplain which has spread to alluvial depression until 1985. During the entire study period (1954–2020), the marshes on the high floodplain (29.585 km²) in wetland reserves were almost uncultivated which only decreased by 12%. Unlike the overall shrinking trend in the region, the marshes distributed on the micro-knoll (24.531 km²) were quickly reclaimed after 1985, only 2% of the initial area was left until 2005 and completely disappear in 2020. The marshes distributed on alluvial plain (254.557 km²) was the largest which was degraded most seriously. Since 1985, large-scale and rapid farmland opening had occurred on alluvial plain, and the result showed that the marshes in this area were only 2.54 km² in the end. The area of marshes distributed on depressions were 128.86 km² and 104.46 km², respectively. The time node of lower terraces was basically the same as alluvial plain which was only 3% were retained. The marshes on depression had been reclaimed with highly speed after 2000, only 4% of the initial area were retained in 2020 (Fig. 8, Table 6).

Discussion
The historical stage of the change of marshes. In the period of before establishment of ‘new’ China, the sub-Sanjiang Plain had just experienced the ban policy of the Qing government, which had the sparsely populated and intact marshes accounting for 94% of the total area of the transect. However, we found that only 0.09% of the total area of marshes in sub-Sangjiang Plain was remain by 2020, the landscape pattern of which had changed completely. Agricultural encroachment caused 86% of the total reduction in marshes in sun-Sangjiang Plain during the study period.
Plain from 1954 to 2020 (Fig. 9). Which was conformity with previous conclusions that agriculture activities are the most main driving cause of marshes7,27,49.

Throughout the shrinking process of the marshes in the sub-Sanjiang Plain, although the area of the marshes in the sub-Sanjiang Plain continued to decrease from 1954 to 2020, the whole process had obvious historical stages characteristic (Fig. 3, Table 4). The historical stage characteristics of marshes degradation in the sub-Sanjiang Plain reflect the national government’s land policy, demand, and scientific and technological progress after the founding of the People’s Republic of China29.

Before 2005, the annual degradation rate of marshes in sub-Sangjiang Plain continued increase. The annual degradation rate of marshes from only 0.2% in the reform and opening up (1954–1976) to 6.29% in the period of transformation reform and opening up (1995–2005), increased 35 times approximately. But which was not a uniform growth and had its own historical characteristics. The “Great Leap Forward” (1958–1960) and “The Down to the Countryside Movement” (1970–1972) activities, made the marshes begun to be reclaimed27,50 by a large number of immigrants poured into the Sanjiang Plain. Then the land reclamation demand of “Modern Farm Construction” (1979–1983) and the stimulation of the household responsibility system51 both promote the enthusiasm for reclamation. However, due to the limitation of technical conditions the degree of degradation was not serious at begin. With the support of the “Japan Loan Project” many large farming machinery and equipment had been put into use the rate and area of reclamation rose rapidly52, the rate of reclamation began to rise rapidly. Why the annual degradation rate in the period of transformation reform and opening up (1995–2005) reached 6.29%, the biggest? That’s because of nearly 40 years of water conservancy project construction, Sangjiang Plain had initially formed three major engineering systems for flood control, waterlogging removal, and irrigation53. Good water conservancy and technology conditions provided the greatest convenience for the rapid development of agriculture.

Although the annual degradation rate of marshes was still high, it had begun to decline after 2005. That’s all because of the never slackened protection of wetlands since China joined the “Convention on Wetlands of International Importance, Especially as Waterfowl Habitats” in 1992. In 2002, China implemented the “National Wetland Conservation Plan” (NWCP) and issued two plans for the pre-wetland protection project. In 2014, the red line of the wetland nature reserve was delineated. A series of policies about protection wetlands achieved significant effects in the protection and restoration, especially the wetlands in the Sanjiang Plain54, in addition, a national plan to adjust agricultural plantation structures was promoted and implemented actively by the local government, large areas of dry land converted into higher economic benefits paddy fields, reducing the reclamation of marshes. However, the policies of direct agricultural subsidy on crops after 2004 and cancelling taxes on crops from 2006 might have also encouraged farmers to convert marshes to cropland to get more income38,55,56. Additionally, the establishment of the World Trade Organization (WTO) in 2001 expanded the grain trading market which also made grain prices rose16. Then serious illegal cultivation has occurred. In a word, the drive of benefits caused the loss of marshes and the continuous increase of farmland.

The spatial–temporal characteristics of the change of marshes. The statistical results of the first regions and sub-regions in geomorphology indicated that the degradation process of marshes in the sub-Sanjiang Plain showed obvious spatial location differences in different historical stages and the degree of reclaiming of marshes on different geomorphic was different. From 1954 to 2020, because of the geological tectonic, the marshes degradation first occurred in the meco-scale recent slow ascent region which was most serious region, and then expanded to the substantially recent slow subsidence region and the small-amplitude recent slow ascent region. It was worth to note that the marshes degradation in the small-amplitude recent slow ascent region was still a trend to continue to expand. In different geomorphic regions, the degradation of marshes first began in alluvial plain, gradually spread to lower terrace and high floodplain, then followed to micro-knoll and low floodplain, and finally extended to depressions. This result fully reflected the restriction on farmland recla-
mation by geomorphic conditions, reflecting the evolution of the man-land interrelations in the sub-Sanjiang Plain during 1954–2020.

Before the period of comprehensive reform and opening up (1985–1995), marshes reclamation mainly relied on manpower and nature conditions while being restricted by nature in the period of relative backward time which lack of machinery and equipment. Therefore, the reclamation of farmland first occurred in places with superior natural conditions, convenient transportation and easy access, and high population density. And the expansion of reclamation was centered on newly-built residential areas.

As Figs. 3 and 5 shows that both region I and III were sporadic reclamation, continuous reclamation only in region II, and all occurred on the more accessible geomorphic alluvial plains, lower terrace and floodplains (Table 6). Since then, the hydrological cycle of marshes had been changed because of the technology and productivity improved, reconstruction of wells, channels and well irrigation marshes project and continuous farmland reclamation which caused the edges of marshes continue to retreat to the depths of the marshes, and the boundaries of farmland continued to invaded. Conversions between marshes and farmland set up a positive feedback system in the marshes which transforms the marshes into upland, suitable for agricultural operations, and increase of agricultural operations further transforms parts of the marshes. So far, human beings have moved from the stage of being greatly restricted by nature into a process in which humans can vigorously transform nature.

The reclamation sites of marshes had expanded from relatively flat alluvial plains and lower terrace to unsuitable micro-knoll and depressions. The distribution of farmland has also begun to deviate from residential areas and expand linearly along ditches and rivers. In the period of deepening reform and opening up (2005–2020), due to the insufficient reserve resources and implementing national protective measures, the degradation process of marshes had been suppressed to a certain extent, then the man-land interrelations in the sub-Sanjiang Plain had begun to develop in a good direction.

**Implications of natural wetland management.** The degradation of marshes in sub-Sanjiang Plain had been controlled to a certain extent, however, the marshes in sub-Sanjiang Plain still urgently need targeted management measures to promote the sustainable development of marshes.

Our study found that most of the existing wetlands in the sub-Sanjiang Plain exist in the National Nature Reserve. Which shows that the establishment of nature reserves plays an important role in the protection of wetlands. Therefore, the conservation and rehabilitation of natural wetlands needs consistent attention from both the government and the public. First, increasing the effectiveness protected areas of wetlands are essential measures for wetlands management. Second, improve the supervision of government to decline the phenomenon of illegal reclamation. Third, a sound national regulatory law is needed to solve the problem of varied protection level in natural wetland conservations several caused by government departments jointly controlled.

Research results proved that the degradation degree and sequence of marshes distributed on different geomorphic types are different. So, we reasonably inferred that the wetlands restoration speed and ultimate degree of restoration with different geomorphological backgrounds will also be limited by its natural conditions. Therefore, the priority areas for restoration should thus focus on the wetlands lost recently such as degraded wetlands isolated from large areas (depressions and micro-knoll), or areas surrounding important wetlands for biodiversity conservation and ecosystem services. As for areas that have been cultivated for a long time such as alluvial plains, considering the changes in food security and hydrological conditions, dry farmland can be transformed to green paddy fields to create artificial wetlands. Improving the living environment quality of waterfowl.

It is impossible to separate the wetlands from agriculture on the Sanjiang Plain. Consider only the protection of wetlands without considering the development of agriculture is unadvisable. The serious problems faced by wetlands are not only direct erosion, but also the opposition between food security and wetland protection, water pollution and reduce chemical pollution; change the flood irrigation method of paddy field to centralized irrigation to reduce the conflict between paddy field irrigation and wetland water use.

**Conclusion**

According to the historical phases, this study adopted the time series of marsh dynamic change datasets derived from the existing land use data and Landsat series remote sensing images which revealed the changes of land-use/cover patterns in the sub-sanjiang Plain. Study showed that the main reason for the shrinkage of the marshes was the occupation of farmland and the process of marshes farming is the evolution process man-land interrelations in the sub-sanjiang Plain. The process of marshes farming in the sub-sanjiang Plain exhibited significant historical stages and location characteristics with the development of the China and policy orientation was an important factor in the transformation of the human-land interrelations.

The period of transformation reform and opening up (1995–2005) was the most severe period of marshes degradation in the sub-sanjiang Plain which was also the most intense period of human-land conflict. After 2005, the degradation of marshes was prevented in a certain extent due to various protection policies, measures, and the establishment of natural reserves of wetlands. Most of the existing wetlands are distributed in national natural reserves. The marsh degradation was determined by the general regional geomorphic types.

During the study period, the degradation first occurred in the meco-scale recent slow ascent region, and then extended to the substantially recent slow subsidence region and the small-amplitude recent slow ascent region. The process of marshes degradation sequentially consisted of firstly began in alluvial plain, gradually spread to lower terrace and high floodplain; then followed to micro-knoll and low floodplain, and finally extended to depressions.
Data availability
The datasets used during the current study are available from the corresponding author on reasonable request.

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References
1. Moor, H., Hylander, K. & Norberg, J. Predicting climate change effects on wetland ecosystem services using species distribution modeling and plant functional traits. *Ambio* **44**(1), 113–126 (2015).
2. Moreno-Mateos, D., Power, M. E., Comín, F. A. & Yockteng, R. Structural and functional loss in restored wetland ecosystems. *PLoS Biol.* **10**(1), e1001247 (2012).
3. Huang, N., Wang, Z. M., Liu, D. W. & Zheng, N. Selecting sites for converting farmlands to wetlands in the Sanjizong Plain, Northeast China, based on remote sensing and GIS. *Environ. Manag.* **46**(5), 790–800 (2010).
4. Bai, J. H., Lu, Q. Q., Zhao, Q. Q., Wang, J. J. & Ouyang, H. Effects of alpine wetland landscapes on regional climate on the Zoige Plateau of China. *Adv. Meteorol.* **2013**(5), 1–7 (2013).
5. Junk, W. J. et al. Current state of knowledge regarding the world’s wetlands and their future under global climate change: A synthesis. *Aquat. Sci.* **75**(1), 151–167 (2013).
6. Gardner, R. C. & Finlayson, M. Global wetland outlook: State of the world’s wetlands and their services to people 2018. Secretariat of the Ramsar Convention (2018).
7. Mao, D. H. et al. China’s wetlands loss to urban expansion. *Land Degrad. Dev.* **29**(8), 2644–2657 (2018).
8. Song, K. S. et al. Wetlands shrinkage, fragmentation and their links to agriculture in the Muleng-Xingkai Plain, China. *J. Environ. Manag.* **111**, 120–132 (2012).
9. Tomaselli, V., Tenerelli, P. & Scandrello, S. Mapping and quantifying habitat fragmentation in small coastal areas: A case study of three protected wetlands in Apulia (Italy). *Environ. Monit. Assess.* **184**(2), 603–713 (2012).
10. Saha, T. K. & Pal, S. Emerging conflict between agriculture extension and physical existence of wetland in post-dam period in Atreyee River basin of Indo-Bangladesh. *Environ. Dev. Sustain.* **21**(3), 1458–1505 (2019).
11. Bayley, S. E., Wong, A. S. & Thompson, J. E. Effects of agricultural encroachment and drought on wetlands and shallow lakes in the boreal transition zone of eniin. *Wetlands** **33**(1), 17–28 (2013).
12. Furlan, L. M., Moreira, C. A., Alencar, P. G. & Rosolen, V. Environmental monitoring and hydrological simulations of a natural wetland based on high-resolution unmanned aerial vehicle data (Paulista Peripheral Depression, Brazil). *Environ. Challenge* **4**, 100146–100151 (2021).
13. Rebelo, L. M., Finlayson, C. M. & Nagabhattachary, N. Remote sensing and GIS for wetland inventory mapping and change analysis. *J. Environ. Manag.* **90**(7), 2174–2153 (2009).
14. Reis, V. et al. A global assessment of inland wetland conservation status. *Bioscience* **67**(6), 523–533 (2017).
15. Mahafei, A., Diwediga, B., Fousseni, F., Wala, K. & Akpagana, K. Landscape-based analysis of wetlands patterns in the Ogu River basin in Togo (West Africa). *Environ. Challenge* **2**, 100013 (2021).
16. Mao, D. H. et al. Conversions between natural wetlands and farmland in China: A multiscale geospatial analysis. *Sci. Total Environ.* **634**, 550–560 (2018).
17. Richards, D. R. & Friess, D. A. Rates and drivers of mangrove deforestation in Southeast Asia, 2000–2012. *Proc. Natl. Acad. Sci.* **113**(2), 344–349 (2016).
18. Fu, B. J. et al. Mechanisms of human-natural system coupling and optimization of the Yellow River Basin. *China Sci. Found.* **35**(04), 504–509 (2021).
19. Wang, M. G. et al. Separating the effects of climate change and human activities on drought propagation via a natural and human-impacted catchment comparison method. *J. Hydrol.* **603**, 126913 (2021).
20. Niu, H. G., Song, H. Y. & Meng, X. M. The marsh genesis and the law of peat distribution in the Sanjiang Plain. *Sci. Geogr. Sin.* **10**(003), 246–256 (1990).
21. Lisenby, P. E., Tooth, S. & Ralph, T. J. Product vs. process? The role of geomorphology in wetland characterization. *Sci. Total Environ.* **663**, 980–991 (2019).
22. Yang, X. D., Shan, G. J., Wang, L. Y. & Xiao, H. J. Physiognomy factor and main environment geology problems of Sanjiang Plain. *Environ. Sci. Manag.* **30**(04), 74–78 (2005).
23. Chen, X. B. Genesis of marsh in Sanjiang Plain. *Ref. Middle Sch. Geogr. Teach.* **7**, 607 (1993).
24. Mitchell, M., Herman, J., Bilikovic, D. M. & Hershner, C. Marsh persistence under sea-level rise is controlled by multiple, geologically variable stressors. *Ecosyst. Health Sustain.* **3**(10), 1379988 (2017).
25. Singh, M. & Sinha, R. Distribution, diversity, and geomorphic evolution of floodplain wetlands and wetland complexes in the Ganga plains of north Bihar, India. *Geomorphology* **351**, 106960 (2019).
26. Phillips, J. D. Coastal wetlands, sea level, and the dimensions of geomorphic resilience. *Geomorphology* **305**, 173–184 (2017).
27. Wang, Z. M. et al. Process of land conversion from Marsh into cropland in the Sanjiang Plain during 1954–2005. *Wetl. Sci.* **7**(3), 208–217 (2009).
28. Wang, Z. M. et al. Changes of land use and of ecosystem service values in Sanjiang Plain, Northeast China. *Environ. Monit. Assess.* **112**(1–3), 69–91 (2006).
29. Wang, Z. M. et al. Loss and fragmentation of marshes in the Sanjiang Plain, Northeast China, 1954–2005. *Wetlands* **31**(5), 945–954 (2011).
30. Liu, X. H. et al. Characterizing the spatial pattern of marshlands in the Sanjiang Plain, Northeast China. *Ecol. Eng.* **53**, 335–342 (2013).
31. Zhang, S. Q. et al. Identifying wetland change in China’s Sanjiang Plain using remote sensing. *Wetlands* **29**(1), 302–313 (2009).
32. He, Y. & Zhang, M. X. Study on wetland loss and its reasons in China. *Chin. Geogr. Sci.* **11**(3), 241–245 (2001).
33. Xu, Y. M. et al. The impact of long-term agricultural development on the wetlands landscape pattern in Sanjiang Plain. *Procedia Environ.* **13**, 1922–1932 (2012).
34. Yue, S. F., Yan, Y. C., Zhang, S. W., Yang, J. C. & Wang, W. J. Spatiotemporal variations of soil freeze-thaw state in Northeast China based on the ERA5-LAND dataset. *Aquat. Geogr. Sci.* **76**(11), 2765–2779 (2021).
35. Shu, L. C. et al. Land use change and its impact on groundwater in the typical district of Sanjiang Plain during the past 40 years. *I. Hydraul. Eng.* **52**(08), 896–906 (2021).
36. Meng, H. Research on the Impact of Climate Change on the Marsh Distribution and Its Risk Assessment in the Sanjiang Plain (University of Chinese Academy of Sciences, Northeast Institute of Geography and Agroecology, Chinese Academy of Science, 2016).
37. Liu, H. Y., Lv, X. G. & Xiang, S. K. Landscape biodiversity of wetlands and their changes in 50 years in watersheds of the Sanjiang Plain. *Acta Ecol. Sin.* **24**(7), 1472–1479 (2004).
38. Yang, C. X., Zheng, H. & OuYang, Z. Y. Changes, effect and driving forces of land use in Sanjiang Plain. *Environ. Prot. Sci.* **221**(05), 103–108 (2020).
39. Ge, Y. N., Zhang, Q. H., Dong, X. H. & Yang, X. D. Revealing anthropogenic effects on lakes and wetlands: Pollen-based environmental changes of Liangzi Lake, China over the last 150 years. *CATENA* **207**, 105605 (2021).
40. Mao, D. H. *et al.* Wetland changes in the Amur River Basin: Differing trends and proximate causes on the Chinese and Russian sides. *J. Environ. Manag.* **280**, 111670–111671 (2021).
41. Yang, L. H. Open governance: The most valuable experience of national governance in China’s 40 years of reform and opening up. *J. Taiyuan Univ. Technol. (Soc. Sci. Ed.)* **39**(06), 1–11 (2021).
42. Liu, Y., Liu, Y. S., Guo, L. Y. & Lu, S. S. Spatial-temporal patterns of land-use/cover change in typical transect area along China National Highway 106 during 1996–2008. *J. Northeast Agric. Univ. (Engl. Ed.)* **18**(3), 39–46 (2011).
43. Koch, G. W., Scholes, R. J., Steffen, W. L., Vitousek, P. M. & Walker, B. H. The IGBP terrestrial transects: Science plan. Global Change Report (Sweden) (1995).
44. Liu, W. *et al.* Spatiotemporal patterns and drivers of methane uptake across a climate transect in Inner Mongolia Stepppe. *Sci. Total Environ.* **757**, 143768 (2021).
45. Zhang, X. S. & Yang, D. Y. Allocation and study on global change transects in China. *Quat. Sci.* **15**(1), 43–52 (1995).
46. Gao, Y., Su, F., Sun X. Y., Xue, Z. S. & He, Y. W. A study on spatial and temporal variations of coastal wetland in Pearl River Estuary. In *2010 IEEE International Geoscience and Remote Sensing Symposium (IGARSS)*, 429–432 (IEEE, 2010).
47. Liu, Y. *et al.* Spatial patterns and driving forces of land use change in China during the early 21st century. *J. Geogr. Sci.* **20**(4), 483–494 (2010).
48. Jia, W. H. *Summarization on China Dynasty’s Policy of Blockage in Northeast China* (Northeast Normal University, 2007).
49. Zhang, W. Q. & Song, G. Study on the land use change and effect factors in the process of land conversion from wetland into cultivated land in Sangjiang plain. *Northeast Horticult.* **19**, 144–152 (2019).
50. Song, K. S. *et al.* Land use change in Sanjiang plain and its driving forces analysis since 1954. *Acta Geogr. Sin.* **63**(1), 93–104 (2008).
51. Bryan, B. A. *et al.* China’s response to a national land-system sustainability emergency. *Nature* **559**(7713), 193–204 (2018).
52. Song, K. S. *et al.* Wetland degradation: Its driving forces and environmental impacts in the Sanjiang Plain, China. *Environ. Manag.* **54**(2), 255–271 (2014).
53. Li, X. L., Han, J. C., Wang, S. L. & Liu, M. X. Analysis of current situation and main problems of water conservancy project of Jian Sanjiang Branch. *Mod. Agric.* **000**(001), 17–18 (2008).
54. Xiang, H. X. *et al.* What did China’s National Wetland Conservation Program Achieve? Observations of changes in land cover and ecosystem services in the Sanjiang Plain. *J. Environ. Manag.* **267**, 110623 (2020).
55. Zhang, C. L., Tong, L. J. & Liu, J. B. Evaluation of coordinated development of arable land and wetlands in Sanjiang reserve. *Sci. Geogr. Sin.* **28**, 343–347 (2008).
56. Zhang, H. Q. & Ma, F. C. Fluctuation analysis and evaluation of grain production in Heilongjiang province. *J. Heilongjiang Bayi Agric. Univ.* **000**(003), 97–101 (2013).
57. Yu, X. F., Sun, M. Y. & Xue, Z. S. Wetland recreational agriculture: Balancing wetland conservation and agro-development. *Environ. Sci. Policy* **87**, 11–17 (2015).
58. Liu, Y., Peng, L. & Chen, Y. Conversion of land use types and ecological effect in Northeast China. *Trans. Chin. Soc. Agric. Eng.* **21**(11), 175–178 (2015).
59. Zou, Y. C. *et al.* Water use conflict between wetland and agriculture. *J. Environ. Manag.* **224**, 140–146 (2018).

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Author contributions
J.T. and Y.L., conceived and coordinated the research. J.T., B.F. and X.J. contributed to the research design. J.T., X.J. and X.Z. collected the data and conducted the analysis. Y.L. contributed to the data sources. Y.G., Y.L. and J.T. contributed to the manuscript and edited the manuscript.

Competing interests
The authors declare no competing interests.

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