Full length article

Exposure to naphthalene and β-pinene-derived secondary organic aerosol induced divergent changes in transcript levels of BEAS-2B cells

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ABSTRACT

The health effects of exposure to secondary organic aerosols (SOAs) are still limited. Here, we investigated and compared the toxicities of soot particles (SP) coated with β-pinene SOA (SOA β-pinene-SP) and SP coated with naphthalene SOA (SOA Naph-SP) in a human bronchial epithelial cell line (BEAS-2B) residing at the air–liquid interface. SOA β-pinene-SP mostly contained oxygenated aliphatic compounds from β-pinene photooxidation, whereas SOA Naph-SP contained a significant fraction of oxygenated aromatic products under similar conditions. Following exposure, genome-wide transcriptome responses showed an Nrf2 oxidative stress response, particularly for SOA Naph-SP. Other signaling pathways, such as redox signaling, inflammatory signaling, and the involvement of...
1. Introduction

Massive urbanization and industrialization have increased particulate matter (PM) levels in the atmosphere, generating high loading of mixed organic/inorganic particles, including soot particles (SP), dust, and smoke from biomass burning (Brauer et al., 2016; Manisalidis et al., 2020; Pöschl and Shiraiwa, 2015; Schnitzler et al., 2014). A dominant component of atmospheric PM is secondary organic aerosols (SOAs), which are formed through the oxidation of volatile organic compounds (VOCs) from biogenic and anthropogenic sources (Nault et al., 2021). More biogenic VOCs are emitted to the atmosphere globally than typical anthropogenic precursor have higher toxicological potency, which was accompanied with the activation of varied cellular mechanisms, such as aryl hydrocarbon receptor. This can be attributed to the difference in chemical composition; specifically, the aromatic compounds in the naphthalene-derived SOA had higher cytotoxic potential than that of the β-pinene-derived SOA.

Recent technological advances enable the use of RNA sequencing (RNA-seq) to uncover multiple facets at the transcript level. RNA-seq allows the relative changes in each transcript to be quantified under specific treatment conditions or during defined developmental stages (Han et al., 2015; Koch et al., 2018). RNA-seq is not limited to selected probes on an array and can be applied to the whole genome for expression patterns that have not yet been established (Han et al., 2015). Therefore, RNA-seq may be an appropriate tool to understand biological complexity and facilitate biologically relevant changes in SOA toxicity.

This study investigates genome-wide activation pathways of exposure to low levels of SP with a low content of organic matter and SP coated by SOA (SOA-SP) on BEAS-2B lung cells (non-tumorigenic lung cells). For this purpose, two chemically different SOA precursors that represent biogenic or anthropogenic sources with different chemical characteristics were chosen. Important biogenic compounds involved in aerosol formation are monoterpenes (Alves and Pio, 2005; Mehra et al., 2020; Ylisirniö et al., 2020) and aromatics (Kolb and Worsnop, 2012; Kroll and Seinfeld 2008; Lim et al., 2019). SOA composition is convoluted and consists of a large variety of complex organic compounds (Nault et al., 2021). These secondary compounds either nucleate and form new particles or condense onto existing particles such as soot or dust, creating complex organic matrix coating (Schnitzler et al., 2014). Therefore, understanding the relationship between the toxicity of SOA-coated particles, their chemical composition, and how atmospheric aging modifies the toxicity is a key to determine their effects on the climate and human health.

Exposure to ambient PM has been associated with significant adverse health outcomes that in part account for high morbidity, and mortality worldwide (Burnett et al., 2018; Cohen et al., 2017; Forouzanfar et al., 2016; Manisalidis et al., 2020; Nault et al., 2021). As a result, the World Health Organization (WHO) has recently updated its guidelines, reducing PM recommendations even more (World Health 2021). Toxicology studies suggest that the possible mechanisms for PM-induced health effects are through PM-induced oxidants production (such as reactive oxygen and nitrogen species ROS/RNS) and oxidative stress, inflammatory response, cellular damage, and death (Al Housseiny et al., 2020; Bai et al., 2017; Burnett et al., 2018; De Grove et al., 2018; Longhin et al., 2018; Wang et al., 2017; Zheng et al., 2017). To date, several studies have examined the toxicity induced by exposure to SOA (Khan et al., 2021; Lund et al., 2013; Tuet et al., 2017a; Tuet et al., 2017b). We have previously shown that exposure to SP coated with naphthalene SOA leads to a stronger toxic response compared to that of SP coated with β-pinene SOA in A549 cells and in a coculture model containing A549 cells and EA.hy926 hybrid human endothelial cells (Offer et al., 2022). It was also shown that aged SOAs from α-pinene and naphthalene in the absence of NOx contained more ROS than fresh SOA and induced a strong cell death response in A549 cells (Chowdhury et al., 2019). Furthermore, at a high dose of 10 μg/mL, fresh extracts from α-pinene, m-xylene, and trimethylbenzene SOA increased the expression of the hemeoxygenase-1 (HMOX-1) gene in both bronchial epithelial BEAS-2B cells and macrophage-like U937 cells (Ito et al., 2019). Taken together, these studies highlight the role of chemical composition in determining the toxicities of SOA. While there have been several recent studies regarding the health effects of SOA (Chowdhury et al., 2019; Chowdhury et al., 2018; Han et al., 2020; Ito et al., 2019; Khan et al., 2021; Tuet et al., 2017a), there are still important knowledge gaps in comparison with those of primary aerosols, but also in comparison between anthropogenic and biogenic toxicities, that have not yet been addressed.

2. Methods

2.1. Generation of SP, SOAβPin-SP, or SOANap-SP

SOAs of naphthalene (Sigma-Aldrich, 99%, GE) or β-pinene (Sigma- Aldrich, ≥ 99%, GE) were produced by OH-dominated photooxidation in a potential aerosol mass (PAM) (Kang et al., 2007; Lambe et al., 2011) oxidation flow reactor (OFR), consisting of a metal cylinder and operated in OFR185 mode (Bruns et al., 2015; Lambe et al., 2011). The SOA condensed onto SP, which are low in organic carbon, from a Combustion Aerosol STandard (CAST) soot generator (Jing Inc) with propane fuel. Fresh SP were used as a reference control. Detailed information regarding the aerosol generation is described in Offer et al. (2022).

2.2. Physical and chemical characterization of SP, SOAβPin-SP or SOANap-SP

Particle number and particle size distributions were evaluated using a scanning mobility particle sizer (SMPS) (Condensation particle counter (CPC), TSI Model 3775). Several types of chemical analysis, different mobility analyzer, TSI Model 3082, USA). Equivalent Black Carbon (BC) mass concentration was measured by a 7-wavelength Aethalometer® (Model Magee AE33, Aerosol d.o.o., Slovenia). The photochemical age of the SOA was
evaluated using co-injected deuterated butanol (D9-butanol), which was used as a "photochemical clock" (Barmet et al., 2012) and probed with a quadrupole proton-transfer-reaction mass spectrometer (PTR-MS, Ionicon, Austria). Elemental ratios were measured with a high-resolution time-of-flight aerosol mass spectrometer (AMS, Aerodyne Inc., Billerica, MA, USA). ROS measurements were conducted using an online instrument for particle-bound ROS (OPROSI) (Wragg et al., 2016). Particulate carbon was evaluated by a thermal-optical carbon analyzer (Desert Research Institute Model 2001A, Atmospheric Inc., Calabasas, CA, USA). Transmission electron microscopy (TEM) images were taken by TEM (JEM-2100F, JEOL Ltd, JP). Chemical SOA composition from filter samples was investigated by direct insertion probe – high-resolution time-of-flight MS (DIP-HRTOFMS; Pegasus® GC-HRT 4D; LECO, St. Joseph, MI, USA) (Käfer et al., 2019), and two-dimensional gas chromatography time-of-flight mass spectrometry (GC × GC-TOFMS; Pegasus® BT 4D GC × GC, LECO, St. Joseph, MI, USA). Particulate carbon was evaluated by a thermal-optical carbon analyzer (Desert Research Institute Model 2001A, Atmospheric Inc., Calabasas, CA, USA). All methods are described in detail in (Offer et al., 2022) and the SI. Additionally, methanolic extracts of SP-SOA were examined by ultrahigh-resolution Fourier-transform ion cyclotron resonance mass spectrometry (Solarix, 7 T FT-ICR MS, Bruker Daltoniks, Bremen, GE) with electrospray ionization in positive and negative mode as described in the SI.

2.3. Cell culture and exposures

The BEAS-2B cell line, an SV-40-transformed human bronchial epithelial cell line (obtained from ATCC, No. CRL 9609), was cultured with BEBM medium along with all the additives (Lonza/Clonetics Corporation, Cologne, GE) except GA-1000 (gentamycin-amphotericin B mixture), which was replaced with 100 U ml⁻¹ penicillin and 100 µg ml⁻¹ streptomycin (P/S, Sigma-Aldrich, GE). Twenty-four mm transwell inserts (0.4 µm pore-size, Corning, NY, USA) precoated with 0.03 mg ml⁻¹ bovine collagen Type 1 (Gibco, Dublin, IRL) and 0.01 mg ml⁻¹ bovine serum albumin (Sigma-Aldrich, GE) were seeded 4 days prior to the experiments and were kept at 37 °C in humidified air containing 5% CO₂.

Cells were exposed in the ALI system (Vitrocell® Automated Exposure Station Standard Version) with complete BEBM medium supplemented with 15 mM N-2-hydroxyethylpiperazine-N-2-ethane sulfonic acid (HEPES) buffer solution (Thermo Fisher Scientific, GE) in the basolateral compartment (Offer et al., 2022). Cells were then exposed for 4 h to the conditioned (85% r.h. 37 °C) undiluted (1:1), and differentially diluted aerosols (1:3 and 1:100). These dilutions correspond to ≈ 5.6 and 0.56 ng cm⁻², respectively, for SOAFRING-SP and 9.3 and 0.93 ng cm⁻², respectively, for SOANAP-SP (Table S1). All experiments represent a single exposure under certain conditions, repeated for at least three independent times. After exposure, the effects of the aerosols on the cells were examined with several assays, and the exposure medium was collected on ice for direct analysis or frozen at −80 °C for further analysis.

2.4. Cell viability analysis

The PrestoBlue assay was used to measure cell viability, was performed according to the manufacturer’s instructions and is further detailed in the SI section.

2.5. RNA extraction, library construction, sequencing and gene expression

Following exposure, the membranes containing the cells were incubated in RNPreact Cell Reagent (QIAGEN, Hilden, GE) overnight at 4°C. Then, the membranes were removed, and the cells were stored at −20 °C until the RNA extraction. Total RNA was extracted from the cells using an RNA Plus mini kit (QIAGEN, Hilden, GE) according to the manufacturer’s instructions. The quality of the extracted RNA was assessed with a Nanodrop and TapeStation (Agilent Technologies, CA, USA) according to the manufacturer’s instructions. Replicates with high RNA integrity (RIN > 9.8) were processed for RNA-Seq at the Crown Genomics Institute of the Nancy and Stephen Grand Israel National Center for Personalized Medicine, Weizmann Institute of Science. Total RNA (500 ng for each sample) was processed using the Inhouse poly A-based RNA-seq protocol (INCPM mRNA Seq). Libraries were evaluated by Qubit (Thermo Fisher Scientific) and TapeStation (Agilent Technologies, CA, USA). Sequencing libraries were constructed with barcodes to allow 45 samples to be multiplexed on two lanes of an Illumina NovaSeq machine using the SP (100 cycles) protocol. The output was ~ 22.9 million single-end 100-bp reads per sample. Fastq files for each sample were generated by the usage bc1f2fastq v2.20.0.422.

Gene expression was performed on the extracted RNA using a StepOnePlus Reverse Transcription (RT) PCR instrument (Applied Biosystems, Foster City, CA, USA) as previously described (Pardo et al., 2017) and is further detailed in the SI. Primers were purchased from Sigma-Aldrich, IL) and are described in Table S2.

2.6. Bioinformatics analysis

A detailed description of the RNA-seq data analysis and software is provided in the SI section. Differentially expressed genes were determined by a p-adj of < 0.05 and absolute fold changes > 2 and max raw counts > 30. Four replicates were tested for each exposure type. All clean air (CA) samples were grouped together and used for 8 pairwise comparisons against each SP and SOA-SP dilution. Principal component analysis (PCA) and hierarchical clustering (distance: Pearson’s dissimilarity, method: Ward.d) were performed based on the 1,000 most variable genes. CA outlier samples E34A2, E35A1 and E37A2 were excluded from the analysis, as these samples were distinctly different from all other CA samples. Unsupervised analysis was executed to explore a pattern of gene expression by clustering the 1,316 genes that were determined to be differentially expressed genes (DEGs). Standardized, log 2 normalized counts were used for the K-Means clustering analysis that was performed with Rstudio v3.6.1.

To determine the most significantly relevant biological functions and pathways, DEGs were analyzed with Ingenuity Pathways Analysis (IPA; QIAGEN, Hilden, GE; Ingenuity® Systems, https://www.ingenuity.com). Fig. S1 shows the pipeline for the analysis.

2.7. Cytokine detection

The Luminex® technique was used to detect the cytokine expression profiles in the samples. Cytokines were measured with the Milliplex Map system (Merck KGaA, Darmstadt, GE) using the MILLIPLEX MAP Human High Sensitivity T Cell Panel-ImmunoLungology Multiplex Assay. Eight cytokines were chosen for analysis: CXCL11/I-TAC, IFNγ, IL-12p70, IL-1β, IL-23, IL-6, IL-8 and TNFα. Analysis was performed according to the manufacturer’s instructions. Twenty-five microliters of the cell medium was used for the analysis. Each cytokine had a calibration curve within the assay; however, the detection range was different for each measured cytokine; CXCL11/I-TAC: 1.46-6000 pg/ml, IFNγ: 0.61–2500 pg/ml, IL-12p70: 0.49–2000 pg/ml, IL-1β: 0.49–2000 pg/ml, IL-23: 7.23–2500 pg/ml, IL-6: 0.18–750 pg/ml, IL-8: 0.31–2500 pg/ml and TNFα:0.43–1750 pg/ml. Cytokine concentrations were determined by fluorescence intensity. Fluorescence data were analyzed with Millipore Milliplex Analyst version 3.4 according to the manufacturer’s recommendations. Statistical analyses were performed with the calculated concentrations of each sample.

2.8. Oxidative stress marker, malondialdehyde (MDA) detection

Malondialdehyde (MDA), a marker of lipid peroxidation, was
measured as previously described (Wu et al., 2017), and a further description is found in the SI section.

2.9. DNA damage, comet assay

DNA damage was assessed by an alkaline comet assay as previously described (Di Bucchianico et al., 2017) and is further detailed in the SI section.

2.10. Statistical analysis

The data for the cell viability, cytokine levels, DNA damage, MDA levels and transcript levels by qPCR are expressed as the means ± standard deviation (SD) representing three independent experiments. Differences between group means were tested by two-way ANOVA with a mixed effect model with a Geisser-Greenhouse correction. Differences were considered significant at a probability level of \( p < 0.05 \) using Tukey’s honestly significant difference (HSD) hypothesis testing. The statistical analysis and the generation of the graphs were performed with GraphPad #9 software (GraphPad Software La Jolla, CA 92037, USA).

3. Results

3.1. Physical and chemical characterization of SOA\(_{\alpha\beta\text{-Pin}}\)-SP and SOA\(_{\alpha\beta\text{-Nap}}\)-SP particles

Both aerosol particles were generated under similar OH exposure (SOA\(_{\alpha\beta\text{-Nap}}\)-SP: \( 2.4 \times 10^{11} \) s cm\(^{-3} \), SOA\(_{\alpha\beta\text{-Pin}}\)-SP: \( 2.5 \times 10^{11} \) s cm\(^{-3} \)) conditions, equivalent to approximately 3 days of photochemical atmospheric aging. Both SOA had similar elemental carbon (EC) and organic carbon (OC) contents, volatility, size distribution, mass and number concentrations as well as particle shapes (Table S1). TEM imaging revealed that the aerosol particles consisted of collapsed fractal SP agglomerates coated by organic matter from condensation of secondary organic material from the photooxidation of the naphthalene and \( \beta \)-pinene precursors in the PAM chamber. The calculated aerosol particle deposition was 1.6-fold higher for SOA\(_{\alpha\beta\text{-Nap}}\)-SP (9.3 ng cm\(^{-2} \)) than for SOA\(_{\alpha\beta\text{-Pin}}\)-SP (5.6 ng cm\(^{-2} \)) (Table S1).

Untargeted mass spectrometric analyses reveal that in contrast to similar physical properties, the two SOA types differ in chemical composition. At a signal-to-noise ratio of > 1000, two-dimensional gas chromatography time-of-flight mass spectrometry (2-D GC TOF MS) with electron ionization detected 798 and 397 peaks for SOA\(_{\alpha\beta\text{-Pin}}\)-SP and SOA\(_{\alpha\beta\text{-Nap}}\)-SP, respectively. SOA\(_{\alpha\beta\text{-Pin}}\)-SP was dominated by aliphatic semi-volatile organic compounds, with nopinone the major oxidation product from \( \beta \)-pinene. The cyclic structures identified in SOA\(_{\alpha\beta\text{-Pin}}\)-SP indicate the presence of bicyclic carbon backbone of \( \beta \)-pinene (e.g. angelicidenoil) or of monocyclic oxygenated cyclohexane (e.g., cyclohexadiene), lactones (e.g., 6-methyl-3,4-dihydro-2H-pyran-2-one) and furans (e.g. 2,3-dihydro-4-(1-methylpropyl)-furan). The photooxidation products in SOA\(_{\alpha\beta\text{-Nap}}\)-SP retained one or both aromatic rings, leading to oxygenated naphthalenes (e.g., dihydroxy-naphthalene), naphthaquinones and aromatic lactones as well as benzenes with oxygenated substituents (e.g., 1,2-benzenedicarboxylicdehyde). In both SOA types, polyunsaturated hydrocarbons and oxygenated hydrocarbons were detected in small amounts. These compounds may be assigned to primary emissions of the combustion aerosol standard (CAST) or evolve from thermal decomposition of nonrefractory SOA constituents (Claffin and Ziemann 2019; Zhao et al., 2020).

Aerosol particle constituents of lower volatility were analyzed online by aerosol mass spectrometry (AMS) and off-line by direct inlet probe mass spectrometry (DIP-MS). The AMS results showed higher O:C and lower H:C for SOA\(_{\alpha\beta\text{-Nap}}\)-SP than for SOA\(_{\alpha\beta\text{-Pin}}\)-SP, which is consistent with the higher oxidative potential of SOA\(_{\alpha\beta\text{-Nap}}\)-SP measured by the Online Particle-bound Reactive Oxygen Species Instrument (OPROSI) (Zhang et al., 2022). Both, AMS and DIP-MS analyses indicate the abundance of oxygenated aerosol particle constituents in the high-molecular weight fraction, based on thermal fragments. The molecular composition of the high-molecular weight fraction was further explored by ultrahigh resolution mass spectrometry (UHR-MS) with electrospray ionization (ESI) in both positive (+ESI) and negative mode (-ESI) and is further detailed in the SI.

The + ESI mass spectrum of SOA\(_{\alpha\beta\text{-Pin}}\)-SP contain high-molecular weight species with up to 20 oxygen atoms and distinct Gaussian-like distributions of peak intensities centered at approximately m/z 410, 610 and 785 (Fig. 1A). Within each distribution, 14 or 16 amu peak distances indicate homolog series with [-CH\(_2\)-] or [-O-] entities, respectively. Such a mass spectrometric pattern has been previously linked to oligomeric structures formed during photooxidation of SOA precursors (Kalberer et al., 2004), particularly dimers from \( \beta \)-pinene photooxidation (Kenseth et al., 2018). The presence of oligomers was also verified by collision-induced dissociation of the primary ions, which retained the overall mass spectrometric pattern (Fig. S2). Hence, the assigned oligomers were not an artifact from ionization. A plot of the carbon number (#C) vs. the average carbon oxidation state (O:C\(_\text{av}\)) (Kroll et al., 2011) highlights the oligomers and reveals major peak intensities that appeared at multiples of 10 carbon atoms, which is equal to the number carbons in \( \beta \)-pinene (C\(_{10}\H\(_{16}\)O\(_2\)), and leveled off at 60 carbon atoms (Fig. 1B). SOA\(_{\alpha\beta\text{-Pin}}\)-SP has an O:C\(_\text{av}\) of ~0.9, which was derived from an O:C of 0.38 and an H:C of 1.66 and does not differ significantly between dimers and hexamers. However, SOAs from monomers of \( \beta \)-pinene have a higher O:C\(_\text{av}\) of ~0.35, which can be observed for compounds with a maximum of 10 carbon atoms. A representative chemical structure based on elemental ratios of SOA\(_{\alpha\beta\text{-Pin}}\)-SP may be given by diisopropyl acid acetate (C\(_{16}\H\(_{22}\)O\(_2\)), which is a known photooxidation product of monoterpens (Claeys et al., 2009), and its oligomers.

SOA\(_{\alpha\beta\text{-Nap}}\)-SP also features Gaussian-like peak intensity distributions in the + ESI mass spectrum around m/z 240 and 420, but with highest intensities for monomers (Fig. 1C), which agrees with findings from Wang et al., (2018). In the #C vs. O:C\(_\text{av}\) space, the SOA\(_{\alpha\beta\text{-Nap}}\)-SP oligomers appear with multiples of 10 carbon atoms end with tetramers at #C of 40 as naphthalene (C\(_{10}\H\(_{12}\)H\(_2\)), and also contain 10 carbon atoms (Fig. 1D). Similar to SOA\(_{\alpha\beta\text{-Pin}}\)-SP, the O:C\(_\text{av}\) of the oligomers of SOA\(_{\alpha\beta\text{-Nap}}\)-SP remain stable, but increase from the dimer (O:C\(_\text{av}\) ~0.18) to the monomer (~0.25). Based on the average elemental ratios (O:C: 0.44; H:C: 1.07), a representative structure of SOA\(_{\alpha\beta\text{-Nap}}\)-SP may be given by ferulic acid or dimethyl phthalate (C\(_{16}\H\(_{12}\)O\(_4\) as well as its oligomers (Claeys et al., 2009; Ng et al., 2011). The exemplary structures for SOA\(_{\alpha\beta\text{-Pin}}\)-SP and SOA\(_{\alpha\beta\text{-Nap}}\)-SP also illustrate the fundamental difference in aromaticity between the two SOA-SPs, which can be generalized by the concept of the aromaticity index (AI; AI\(_\text{aromatic}$$\geq$$ 0.5) (Koch and Dittmar 2006). For SOA\(_{\alpha\beta\text{-Pin}}\)-SP<1% of the total peak intensity belongs to aromatic compounds, which account for 26% of the total peak intensity in SOA\(_{\alpha\beta\text{-Nap}}\)-SP.

The -ESI mass spectra reveal a strong dominance of the dimers for both aerosol types and a high aromaticity for SOA\(_{\alpha\beta\text{-Nap}}\)-SP (32% of total peak intensity from aromatic compounds). More information about -ESI analysis and an overview of chemical aerosol properties is detailed in the SI (Fig. S3 and Table S3). Additional information about the physical and chemical properties of the particles can be found in Offer et al. (2022).

3.2. Cytotoxicity of SP, SOA\(_{\alpha\beta\text{-Pin}}\)-SP or SOA\(_{\alpha\beta\text{-Nap}}\)-SP

The toxicity of SOA-SP was determined by measuring the metabolic activity with PrestoBlue, a resazurin-based solution. Exposure was performed to SOA\(_{\alpha\beta\text{-Pin}}\)-SP or SOA\(_{\alpha\beta\text{-Nap}}\)-SP versus the control groups that were exposed to either uncoated SP or clean air (CA). Exposure of BEAS-2B cells to SOA\(_{\alpha\beta\text{-Nap}}\)-SP led to a significant reduction in metabolic activity, particularly under undiluted (1:1) conditions (40%) and at a dilution of 1:3 (30%). Exposure to SOA\(_{\alpha\beta\text{-Pin}}\)-SP did not reduce the metabolic activity compared to SP for all dilutions tested (reduction of 10–15% for both the SOA\(_{\alpha\beta\text{-Pin}}\)-SP and SP, compared to CA, respectively) (Fig. 2), suggesting an increased toxic response for SOA\(_{\alpha\beta\text{-Nap}}\)-SP compared to SOA\(_{\alpha\beta\text{-Pin}}\)-SP.

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3.3. Genome-wide transcriptional responses to SOA

RNA-Seq was performed to detect differential gene expression in BEAS-2B cells exposed to SOA
\(\beta\)-Pin-SP or SOA Nap-SP versus the control groups. Since the PrestoBlue assay indicated not >40% impairment of cell viability after 4 h of exposure to the 1:30 and 1:3 dilutions, RNA-seq of samples exposed to these two dilutions represent actual transcript changes.

Data on the expression of 16,968 genes were retained for the RNA-Seq analysis. PCA, based on the most variable genes, showed the largest separation between the CA samples and SOA-SP for the two tested dilutions (Fig. 3).

3.3.1. Differential gene expression from SOAs-SP

The comparisons performed were between the SP dilution 1:3 and all CA samples (SP3vsCA), SOA
\(\beta\)Pin-SP dilution 1:3 and all CA samples (SOA
\(\beta\)Pin-SP3vsCA), and between SOA Nap-SP dilutions 1:3 (SOA Nap-SP30vsCA). In addition, we used the same groups for comparison with a dilution of 1:30, namely, SP30vsCA, SOA
\(\beta\)Pin-SP30vsCA and SOA Nap-SP30vsCA. The analysis yielded 671 DEGs for SP3vsCA (479 upregulated and 192 downregulated), 725 DEGs for SOA
\(\beta\)Pin-SP3vsCA (441 upregulated and 284 downregulated) and 986 DEGs for SOA Nap-SP3vsCA (624 upregulated and 362 downregulated). Additionally, 409 DEG for SP30vsCA (322 upregulated and 87 downregulated), 288 DEG for SOA
\(\beta\)Pin-SP30vsCA (231 upregulated and 57 downregulated) and 673 DEGs for SOA Nap-SP30vsCA (491 upregulated and 182 downregulated) (Fig. S4 and Fig. S5). The RNA-seq analysis focused on the differences at a dilution of 1:3 as fewer DEGs were observed for SP and SOA
\(\beta\)Pin-SP than for SOA Nap-SP at dilutions of 1:30 and 1:3, respectively. In addition, the cell viability assay showed little biological response at a dilution of 1:30 that was thereafter supported with other biological assays.

DEG were analyzed by two approaches. The first approach was used to observe changes at the individual gene level, whereas the second approach was used to obtain a comprehensive overview of groups of genes and their relevant altered pathways.
3.3.2. Individual unbiased gene analysis

To assess the transcriptional changes at the gene level, Venn diagrams (Fig. S4) and volcano plots (Fig. S5) were plotted (Koch et al., 2018). The expression of 15 most significant up- and down-regulated individual genes is presented in Table S4. Among the SP, et al., 2018). The expression of 15 most significant up- and down-regulated individual genes, no special trends were identified between the SP, SOA

3.3.3. Pathway analysis

To obtain a comprehensive insight into the biological functions and pathways influenced by exposure to SOA_{\text{βPin}}-SP or SOA_{\text{Nap}}-SP in lung cells, clustering (Fig. S6) and functional annotation of DEGs were performed using Ingenuity Pathway Analysis (IPA). In the identified pathways, differential expression of genes significantly exceeded a p-value in at least one experimental setting and the activating z-score threshold was surpassed to determine the activation or inhibition of pathways.

Using IPA, each SOA-SP was analyzed separately compared to the CA samples, and relevant canonical pathways were identified. SOA_{\text{βPin}}-SP and SOA_{\text{Nap}}-SP show common canonical pathways related to the inflammatory response, stress response and MAPK signaling. A detailed description of the z score canonical pathway, upstream analysis, disease and function information, and validation by real-time PCR (HMOX-1, Bcl2, IL-8, GADD45, AhR, MMP-3 and MMP-10) (Fig. S7) is presented in the SI. Twenty-five canonical pathways are summarized as pie charts based on their known involvement in various cell responses. The canonical pathway groups were defined as i) “Inflammation and immune response” (ii) “Cancer” (iii) “Cell cycle and proliferation” (iv) “Signaling pathway” and (v) “Oxidative stress response”. While some pathways may be related to more than one group or cluster, the same classification was kept throughout the analysis to highlight the general changes in the activated pathways by the exposure to the different particle types (Table S5). The relative contribution of the canonical pathway of human lung cells exposed to SP and SOA-SP showed that a large portion of the pathways were related to the “inflammatory/immune response” (Fig. 3B and Table S5). These pathways largely consist cells exposed to SP and SOA_{\text{βPin}}-SP. The “Signaling pathway”, related to “MAP kinase signaling”,

Fig. 3. Principal component analysis (PCA) plot of RNA-Seq data and canonical pathway grouping after exposure to SP or SOA-SP. (A) Purple color indicates exposure to SP, red symbols indicate exposure to SOA_{\text{βPin}}-SP, green symbols indicate exposure to SOA_{\text{Nap}}-SP, blue symbols and round shapes indicate exposure to clean air (CA). Triangle shapes indicate a dilution of 1:30, and square shapes indicate a dilution of 1:3. (B) The top 25 canonical pathways used for this analysis were grouped according to their function and cell responses; “Inflammation and immune response”, “Cancer”, “Cell cycle and proliferation”, “Signaling pathway” and “Oxidative stress response”, further description about the canonical pathway grouping strategy is detailed in the SI. The pie charts demonstrate the relative contribution of the major groups. IPA filters were set to p-adj of < 0.05, absolute fold changes > 2 and max raw counts > 30 for enriched canonical pathways.
was more prevalent in cells exposed to SOA<sub>Nap</sub>-SP. Exposure to both SOA<sub>βPin</sub>-SP and SOA<sub>Nap</sub>-SP increased the contribution of the “oxidative stress response” in comparison to SP. No substantial difference was observed for the “Cell cycle and proliferation” and “Cancer” pathways among SP, SOA<sub>βPin</sub>-SP and SOA<sub>Nap</sub>-SP exposures.

### 3.4. Expression of proinflammatory cytokines in cells exposed to SOA-SP

Eight cytokines (IL-6, IL-8, IL-1β, TNF-α, IL-12p70, IFNγ, CXCL11/I-TAC, and IL-23) were analyzed from the medium of the exposed cells to assess the induction of inflammatory response by SOA-SP. These cytokines were identified based on the RNA-seq analysis, but have also been shown to increase in response to PM exposure in human bronchial epithelial cells in previous studies (Das et al., 2021; Longhin et al., 2018). Changes in cytokine secretion levels during exposure to different dilutions of SOA-SP were investigated (Fig. 4). Cytokine levels showed dose-dependent response when exposed to SOA-SP compared to that of the controls (SP and CA) for most cytokines tested, further demonstrating the importance of SOA in generating the inflammatory response. The dilution effect was observed with up to a 1:3 dilution, which avoided significant cell death compared to that of the undiluted levels. Lower levels of cytokines expression was observed in SP than SOA-SP. The expression levels of IL-6 and IL-8 were significantly higher in SOA-SP than in SP at a dilution of 1:3. No significant differences between SOA<sub>βPin</sub>-SP and SOA<sub>Nap</sub>-SP were observed for IL-6, IL-8 or IL-1β.

The expression of TNF-α, IL-12, IFN-γ, CXCL11/I-TAC, and IL-23 was significantly higher in cells exposed to SOA<sub>Nap</sub>-SP than in cells exposed to SOA<sub>βPin</sub>-SP at a dilution of 1:3. The secretion of IL-8, IL-1β, IL-12, IFN-γ, and IL-23 significantly increased in a dose-dependent manner when cells were exposed to SOA<sub>βPin</sub>-SP and SOA<sub>Nap</sub>-SP, especially for the SOA<sub>Nap</sub>-SP at a dilution of 1:3.

### 3.5. Exposure to SOA-SP induces oxidative stress and damage in lung cells

Pathway analysis revealed that exposure to SOA<sub>βPin</sub>-SP and SOA<sub>Nap</sub>-SP are associated with expression of genes related to Nrf2-mediated oxidative stress response, suggesting that the oxidative stress response and subsequent DNA damage may be important mechanisms leading to the adverse biological effects induced by exposure to SOA. Some of the genes related to Nrf-2 mediated oxidative stress response were evaluated by real-time PCR after exposure to SOA-SP. Exposure to SOA-SP showed a statistically significant increase in HMOX-1 and Glutamate-cysteine ligase regulatory subunit (GCLM) gene levels compared to that of SP. Whereas NAD(P)H dehydrogenase (quinone 1) (NQO-1) transcript level increased in SOA<sub>Nap</sub>-SP and not SOA<sub>βPin</sub>-SP. No change was evident in

![Fig. 4. Cytokine expression after exposure to SP and SOA-SP. BEAS-2B cells were exposed for 4 h to 1:3 and 1:30 dilutions of SP, SOA<sub>βPin</sub>-SP or SOA<sub>Nap</sub>-SP and CA. Cytokine levels were detected in the basolateral medium. (A) IL-6, (B) IL-8, (C) IL-1β, (D) TNF-α, (E) IL-12, (F) IFN-γ, (G) CXCL11, and (H) IL-23. The data represent n independent experiments with one technical replicate, compared to cells exposed to CA, SP, n = 3, SOA<sub>βPin</sub>-SP, n = 4, SOA<sub>Nap</sub>-SP, n = 4, for all dilutions, and CA, n = 8. Two-way ANOVA was used for the analysis. Means with different letters are significantly different. Differences were considered significant at a probability level of *p < 0.05 using Tukey’s HSD hypothesis testing.](image-url)
the Glutamate—cysteine ligase catalytic subunit (GCLC) and (Fig. 5A-D). To explore the oxidative stress response, MDA levels were measured after exposure to SOA_{βPin}-SP and SOA_{Nap}-SP. After exposure to 1:3 dilution SOA_{Nap}-SP, BEAS-2B cells showed a significant increase in MDA levels. Conversely, no change in MDA levels was recorded following exposure to SOA_{βPin}-SP and SP at any dilution tested (Fig. 5E).

ROS can damage DNA and lead to genotoxicity and DNA instability (Bai et al., 2017). A comet assay was performed to assess DNA lesions and genomic instability following exposure to SOA_{βPin}-SP and SOA_{Nap}-SP. SOA_{Nap}-SP showed a statistically significant increase in the percentage of DNA in the tail at 1:30 and 1:3 dilutions compared to that of the control cells. No change was observed for SOA_{βPin}-SP and SP compared to the control (Fig. 5F). This implies a stronger genotoxicity response is induced by SOA_{Nap}-SP than by SOA_{βPin}-SP.

Activation of oxidative stress signaling by aryl hydrocarbon receptor (AhR) and induction of the related Cytochrome P450 (Cyp) enzymes (Pardo et al., 2020) were evaluated by real-time PCR after exposure to SOA-SP, as AhR can be directly activated by PAH (Pardo et al., 2020). Exposure to SOA_{Nap}-SP showed a statistically significant increase in AhR and cyp1a1 gene levels compared to that of SOA_{βPin}-SP. No change was evident in cyp1b1 and aldehyde dehydrogenase (ALDH) genes in SOA-SP (Fig. 6), suggesting that the SOA_{Nap}-SP obtains Cyp-specific response and is driven by PAH.

4. Discussion

SOA from biogenic and anthropogenic sources account for a substantial fraction of the total PM exposure (Borbon et al., 2013; Chowdhury et al., 2019; Chowdhury et al., 2018; Eaves et al., 2020; Gu et al., 2021; Guenther et al., 1995; Han et al., 2020; Ito et al., 2019; Tuet et al., 2017a), and can induce adverse health effects upon exposure (Chowdhury et al., 2019; Chowdhury et al., 2018; Eaves et al., 2020; Han et al., 2020; Ito et al., 2019; Khan et al., 2021; Tuet et al., 2017a). It was estimated that the mass per surface area from human inhalation representing realistic conditions or worst case exposure conditions are 0.03 ng cm\(^{-2}\)h\(^{-1}\) and 5 ng cm\(^{-2}\)h\(^{-1}\) (Paur et al., 2011). In this study, the calculated particle mass concentration (dilution 1:3) per hour was 1.4 and 2.2 ng cm\(^{-2}\)h\(^{-1}\) for SOA_{βPin}-SP and SOA_{Nap}-SP, respectively, corresponding to average levels of ambient exposure conditions. Thus, the lung cells were exposed to two different SOA types with comparable...
concentrations and similar physical properties, yet, with fundamentally different chemical compositions. The dose in this study was even lower than that in the previous studies using the ALI system, in which lung cells were exposed to 67 ng cm\(^{-2}\) than that in the previous studies using the ALI system, in which lung different chemical compositions. The dose in this study was even lower concentrations and similar physical properties, yet, with fundamentally

Recent studies have focused on the biological effects of PM from various sources and the relation of these effects to adverse health outcomes (Bai et al., 2017; Burnett et al., 2018; Cohen et al., 2017; De Grove et al., 2018; Shi et al., 2021). Due to the high abundance of atmospheric secondary aerosols, several studies have investigated the cytotoxic effects of exposure to biogenic or anthropogenic SOA toward lung cells (Chowdhury et al., 2019; Chowdhury et al., 2018; Eaves et al., 2020; Han et al., 2020; Ito et al., 2019; Tuet et al., 2017a). In the present study, we investigated the cytotoxicity and signaling mechanisms, and identified different pathways induced by SOA exposure from proxies for biogenic (SOA\(_{\beta\text{-pinene}}\)) versus anthropogenic (SOA\(_{\beta\text{-naphthalene}}\)) precursors. The physical properties of these SOAs are rather similar (size distribution, volatility, particle shape etc.), but they differ in their chemical nature: SOA\(_{\beta\text{-pinene}}\) is less oxygenated and contains mostly structural increments of the terpene backbone with acid and carbonyl functional groups, while SOA\(_{\beta\text{-naphthalene}}\) is more oxygenated and often retains aromatic rings, which is reminiscent of the naphthalene precursor. Both SOA types contain oligomers, but SOA\(_{\beta\text{-pinene}}\) contains a larger amount of oligomers that are higher and beyond trimers compared to that of SOA\(_{\beta\text{-naphthalene}}\). Here we discuss the chemical differences between the two SOA types that contribute to differential biological responses following exposure to BEAS-2B cells.

To identify the biological functions and signaling pathways activated by SOA exposure lung cells, we analyzed and compared genome-wide transcriptome responses after exposing BEAS-2B lung epithelial cells to SOA\(_{\beta\text{-pinene}}\) and SOA\(_{\beta\text{-naphthalene}}\). In this study, both SOA\(_{\beta\text{-pinene}}\) and SOA\(_{\beta\text{-naphthalene}}\) induced an increase in the genes related to Nrf2 oxidative stress response, particularly SOA\(_{\beta\text{-naphthalene}}\). In addition, canonical pathway grouping showed that both SOA\(_{\beta\text{-pinene}}\) and SOA\(_{\beta\text{-naphthalene}}\) led to an increased abundance in the "oxidative stress response". The overall profile of the activated biological functions influenced by SOA\(_{\beta\text{-pinene}}\) and SOA\(_{\beta\text{-naphthalene}}\) is consistent with the previously reported responses in BEAS-2B cells. A recent study showed that proteome-wide effects of naphthalene-derived SOA significantly upregulated Nrf2-regulated proteins (e.g., NQO1) in BEAS-2B cells (Han et al., 2020). Other SOA types also alter oxidative stress genes: isoprene-derived SOA was shown to alter microRNA (miRNA) expression related to oxidative stress in BEAS-2B cells (Eaves et al., 2020) as well as the induction of the Nrf2 transcription factor and its related genes (Lin et al., 2017). These findings support the notion that SOA is involved in the oxidative stress response, specifically by inducing Nrf2 signaling.

Multiple biological and signaling transduction pathways can be triggered by oxidative stress and particle toxicity (Ahmed et al., 2019; Zheng et al., 2017). For example, exposure to dimethyl selenide (DMSe) aerosols induced elevated genotoxicity, DNA damage, mitogen-activated protein kinases (MAPK) and p53-mediated stress responses, as well as downregulated cholesterol biosynthesis, glycolysis, and interleukin IL-4/IL-13 signaling (Ahmed et al., 2019). Exposing BEAS-2B to ambient SOA promoted the PI3K-Akt, JAK-STAT and FGF/FGFR/ MAPK/VEGF signaling pathways, which led to apoptosis and sustained angiogenesis and affected cell proliferation, and these are all important hallmarks of cancer (Zheng et al., 2017). In our study, the induction of MAPK/growth factor-regulated extracellular signal-related kinase (ERK) signaling was detected. These MAPKs (ERKs, JNKs, or p38 MAPKs) are
serine-threonine protein kinases that are activated by high ROS levels and play a major role in oxidative stress signal transduction (Son et al., 2011). It was also found that organics from fine particles (OC, PAHs) (Zheng et al., 2017) and AhR activation (Shi et al., 2021) engage and play a major role in oxidative stress signal transduction (Son et al., 2017). These serine-threonine protein kinases are activated by high ROS levels and play a major role in oxidative stress signal transduction (Son et al., 2017). This notion is also supported by this study, in which SOA_{\text{Nap}}-SP exposure is hence attributed to the aromatic moieties identified by the mass spectrometric analysis in this SOA.

Both the individual gene analysis and the pathway analysis showed increase in the transcript levels of matrix metalloproteinase (MMP). These MMPs play critical roles in matrix remodeling and modulation of inflammation, cell signaling and intracellular pathways (Dagouassat et al., 2012). Several previous studies have shown a connection between PM chemical composition and the involvement of different types of MMP (Campen et al., 2010; Dagouassat et al., 2012; Xing et al., 2021); for example, MMP-9 mRNA expression and activity in the aorta were detected in mice exposed to SOA from whole combustion emissions of gasoline and diesel (Campen et al., 2010). It was shown that mice exposed to PM_{2.5} suspension by oropharyngeal aspiration had increased mRNA levels of MMP-2 and MMP-9 in the heart tissue, and PAHs positively correlated with MMP-9 expression (Xing et al., 2021). Our data show that SOA_{\text{Pin}}-SP and SOA_{\text{Nap}}-SP mainly increased MMP-1, MMP-3 and MMP-10, whereas a higher effect was induced by SOA_{\text{Nap}}-SP than SOA_{\text{Pin}}-SP. This finding is very interesting, as to the best of our knowledge, this is the first time that elevation of these subtypes of MMP have been reported with exposure to SOA. MMP-3 and MMP-10 are secreted from the cells as inactive proMMPs, but they are then activated on the cell surface. These MMPs correlate with respiratory pathologies (Cui et al., 2017); MMP-3 may be involved in the development of lung cancer among chronic obstructive pulmonary disease (COPD) subjects (Brzóska et al., 2014), and MMP-10 is associated with airway inflammation (Kuo et al., 2019). Further studies are therefore needed to better understand the implications of MMP activation in health effects associated with exposure to SOA.

4.2. Inflammatory response after exposure to SOA

An increasing amount of evidence points to the involvement of SOA in the inflammatory-related cascade, oxidative stress, and DNA damage (Al Housseiny et al., 2020; De Grove et al., 2018; Ma 2020). Thus, we examined the expression of proinflammatory markers known to be expressed by lung epithelial cells. BEAS-2B cells exposed to uncoated SP did not exhibit significant inflammatory response (Al Housseiny et al., 2020), although, according to the RNA-seq analysis, the SP may induce gene expression related to an inflammatory response, as also indicated by other studies (Niranjan and Thakur 2017). It was suggested that the functional groups on the particle surface are sufficient to induce inflammatory response (Al Housseiny et al., 2020; De Grove et al., 2018; Totlandsdal et al., 2015) as well as a specific component of PM (Totlandsdal et al., 2015; Tuet et al., 2017a). For example, it was demonstrated that organic compounds in particles emitted from diesel engine exhaust (DEPs) were the main drivers for releasing epithelial IL-1γ, IL-6, IL-8 and GM-CSF cytokines (De Grove et al., 2018; Totlandsdal et al., 2015). Furthermore, SP with varying surface composition induced the expression of proinflammatory genes compared to unmodified soot (Al Housseiny et al., 2020). This notion is also supported by this study, in which SOA_{\text{Nap}}-SP, induced a profound inflammatory response, whereas exposure to the SP, without significant particle surface functionalization or organic coating, induced only very minor inflammatory response.

Aromatic compounds such as PAHs are AhR ligands. It was shown that AhR induces the production of inflammatory cytokines and alters T cell differentiation (O’Driscoll and Mezrich 2018; Vogel et al., 2020). It was recently reported that environmentally persistent free radicals formed in combustion processes may activate AhR and induce Th17-related cytokines (IL-1β, IL-22, IL-33, keratinocyte-derived chemokine), leading to Th17 polarization and IL-17-dependent pulmonary inflammation (Jaligama et al., 2018), suggesting a direct link between the activation of AhR and inflammatory responses (O’Driscoll and Mezrich 2018; Vogel et al., 2020). Our RNA-seq results showed a profound IL-17 response, especially after exposure to SOA_{\text{Nap}}-SP. This result is in line with the involvement of PAH in AhR activation that mediates the inflammatory response. As shown by 2-D gas chromatography mass spectrometry (GC MS) and the aromatics index from FT-ICR analysis, SOA_{\text{Nap}}-SP contained a high aromatic content, while SOA_{\text{Pin}}-SP did not contain such moieties.

The acute response to inhaled particles is mainly coordinated by type 1 immune cells and cytokines, such as IFN-γ, IL-2, IL-12, IL-1β, IL-18 and TNF-α (Ma 2020; Totlandsdal et al., 2015; Tuet et al., 2017a). In our study, exposure to SOA_{\text{Nap}}-SP increased INF-γ, IL-12, IL-1β, and TNF-α compared to that of SP and SOA_{\text{Pin}}-SP, suggesting that the early acute inflammatory response to SOA is involved with type 1 inflammation. Nevertheless, although immune cells were not directly tested in the study, genes involved in the immune response were differentially expressed after exposure to SP and SOA-SP in BEAS-2B cells. Among these DE genes are changes in plasminogen activator inhibitor-2 (SERPINB2), which serves as a marker of type 2 inflammation (T-helper type 2, Th2) and is considered the main mechanism of asthma (Singhania et al., 2018; Zhang et al., 2019). This suggests that both types of inflammatory responses are involved in SOA-induced inflammation. SERPINB2 is also considered to be regulated by AhR (Brauze et al., 2019; Zhang et al., 2019); it was shown that benzo[a]pyrene (BaP), a PAH used as a benchmark for carcinogenicity, induced SERPINB2 gene expression in the UT-SCC-34 laryngeal squamous cell carcinoma cell line, which was partially dependent on AhR (Brauze et al., 2019). However, in our study the levels of the SERPINB2 gene were equally upregulated for SP, SOA_{\text{Pin}}-SP and SOA_{\text{Nap}}-SP, supporting an independent mechanism of SERPINB2 activation and AhR.

A crucial connection between cytokine production and ROS formation was previously found, as PM-induced cytokines such as IL-1α, IL-1β, IL-6, IL-8 and GM-CSF that were significantly decreased after treatment with radical scavengers/antioxidants (De Grove et al., 2018; Wang et al., 2017). The interplay between oxidative stress and inflammation can be activated by the signal transducer and activator of transcription 3 (STAT3) and MAPK, which were also expressed in this study and mediate the expression of a variety of genes in response to cellular stimuli (Ahmed et al., 2019; Kagan et al., 2017; Wang et al., 2017; Zheng et al., 2017).

4.3. Oxidative damage after exposure to SOA

PM toxicity has been associated with the chemical composition of the particles and with the ability of PM to generate ROS (Chowdhury et al., 2019; Tong et al., 2018; Verma et al., 2015). In this study, exposure to SOA_{\text{Nap}}-SP led to augmented peroxide levels and a higher abundance of lipid peroxidation products than that of SOA_{\text{Pin}}-SP exposure (Offer et al., 2022). In addition, SOA_{\text{Nap}}-SP induced a higher ROS production and oxidative potential than that of SOA_{\text{Pin}}-SP, as measured by OPRSOI (Zhang et al., 2022). ROS yields (e.g. radicals, H_2O_2) and the oxidative potential of isoprene, β-pinene, and naphthalene-derived SOA, measured in water and surrogate lung fluid increased (Tong et al., 2018). In particular, naphthalene-derived SOA was found to release substantial amounts of O_3 and H_2O_2. It is known that excess amounts of ROS can cause oxidative stress and contribute to the imbalance between ROS production and their elimination through protective mechanisms by antioxidants inducing a broad range of toxicological effects such as inflammation, mutations, and lipid and DNA damage (Ahmed et al., 2019; Alves and Pio 2005; Chowdhury et al., 2019; Chowdhury et al., 2018; Eaves et al., 2020; Han et al., 2020; Ito et al., 2019; Lima de Albuquerque et al., 2021; Tuet et al., 2017a). RNA-seq analysis showed a profound reduction in alcohol dehydrogenase (ADH), especially for SOA_{\text{Nap}}-SP.
SOA_{Nap}-SP. Reduced ADH activity leads to a low concentration of reduced glutathione (GSH), prompting oxidative stress and carcinogenesis in lung cancer cells (Orywal et al., 2020). This supports the conclusion that exposure to SOA_{Nap}-SP induces an oxidative stress response with increased ROS and MDA levels and reduced ADH.

The DNA damage response was also remarkably higher for SOA_{Nap}-SP than for SOA_{pin}-SP, SP and CA. It was previously shown that naphthalene metabolites, metabolites 1,2-hydroxynaphthalene (NQH2), 1,4-NQH2, 1,2-naphthoquinone (1,2-NQ), and 1,4-NQ induced DNA damage in the presence of nicotinamide adenine dinucleotide (NADH) and metal ions (Ohnishi et al., 2018). However, in a study that evaluated the source contribution of various atmospheric organic aerosols, naphthalene-derived SOA was evaluated as one of the most important sources of atmospheric organic aerosol toxicity, and the results showed increased oxidative and inflammatory responses but minor effects on DNA damage (Pushimi et al., 2021). Our results suggest that due to the increased oxidative and inflammatory responses but minor effects on DNA damage, SOA_{Nap}-SP induces an oxidative stress response with increased ROS and MDA levels and reduced ADH.

While this study provides novel information regarding the cellular response to β-pinene and naphthalene-derived SOA, it is not without limitations. β-pinene and naphthalene were used as representative (model) SOA precursors from biogenic and anthropogenic sources. However, these two precursors do not provide sharp discrimination by the apparent aliphatic and aromatic carbon backbone of the SOA precursors, which were partly maintained after aging. For example, aromatic compounds, such as p-cymene, may be released from vegetation and monoterpenes may be released from industrial processes or wood combustion for residential heating. Additionally, variations within the groups of biogenic and anthropogenic emissions may lead to the formation of different SOA. Consequently, several individual primary aerosols must be investigated under a broad range of relevant atmospheric conditions to connect physical and chemical aerosol properties to toxicological responses. Epithelial responses can be different when sample collection is performed over different time periods or locations. Furthermore, epithelial cytokine expression can be influenced by the particle size, as well as the experimental setup, the selected PM dose and epithelial cell culture. Finally, as RNA-Seq and pathway enrichment analysis have enabled the identification of pathways at the transcript level, more functional validation such as protein levels changes, will be necessary to demonstrate the effects on the phenotypes.

5. Conclusions

In the present study, human bronchial epithelial cells (BEAS-2B) were exposed to low levels of fresh SP coated by SOA from biogenic and anthropogenic SOA precursors (SOA_{Pin}-SP and SOA_{Nap}-SP, respectively). The results highlight that substantially lower cytotoxic effects were observed with pure SP than SP coated with SOA. The differences in the chemical characteristics of the two SOAs studied led to different and varied cytotoxic responses. Transcriptomic gene analysis followed by pathway enrichment analysis revealed that major biological pathways influenced by SOA-SP are associated with elevated oxidative stress, inflammation and cytokine signaling, as well as dysregulated metabolic activities, that play crucial roles in airway toxicity. The induction of Nrf2 pathway has been found to be unique for SOA-SP compared to soot coated particles. Furthermore, the involvement of MPP in SOA toxicity is reported here for the first time, providing evidence of an intricate toxicological response to SOA exposure. In addition, anthropogenically derived SOAs can activate different pathways and have higher toxicological potency than that of biogenically derived SOAs, as naphthalene-derived SOAs induced stronger oxidative stress and genotoxic and inflammatory responses than those of β-pinene-derived SOA. Based on the comprehensive aerosol particle analysis we attribute this difference to the chemical composition, in which the aromatic compounds and higher oxidation state of the naphthalene-derived SOA lead to higher cytotoxic potential than that of SOA from β-pinene.

This study has profound environmental significance, as anthropogenic SOAs represent a substantial fraction of the PM mass in populated areas around the world and contribute to poor air quality. The high concentrations of anthropogenic SOAs in cities, which mainly originate from burning for residential use, industrial use and tailpipe emissions, lead to high concentrations of toxic aerosols in and downwind urban centers (Daellenbach et al., 2020; Fushimi et al., 2021; Lim et al., 2019). It is estimated that exposure to outdoor PM_{2.5} has led to >8 million premature deaths (Burnett et al., 2018; Forouzanfar et al., 2016). This estimation is supported by recent global simulations that attribute 340,000 PM_{2.5}-related premature deaths per year to anthropogenic SOAs annually. However, large uncertainties regarding anthropogenic SOA sources still exist (Nault et al., 2021). Our study suggests that SOA is generally toxic even at low exposure doses, and anthropogenic SOA is more toxic than biogenic SOA based on structural differences in the most relevant SOA precursors. This means that anthropogenic SOA sources need to be better constrained, and their emissions and related SOAs need to be monitored and reduced.

Future work will be necessary to elucidate the atmospheric complexity of SOAs, and discover dynamic biological responses as a result of atmospheric aging. To fully assess the environmental health impacts of anthropogenic and biogenically derived SOA, a comprehensive protein analysis and additional biological functions in primary epithelial cells as well as coculture or 3D model systems would be of great interest.
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Appendix A. Supplementary material

Additional experimental details, and methods including experimental set up and results. Supplementary data to this article can be found online at https://doi.org/10.1016/j.envint.2022.107366.

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