**Immuno flow cytometry in marine phytoplankton research**

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SUMMARY: The developments in the combination of flow cytometry and immunology as a tool to identify, count and examine marine phytoplankton cells are reviewed. The concepts of immunology and flow cytometry are described. A distinction is made between quantitative and qualitative immunofluorescence. Quantitative immunofluorescence, the identification and enumeration of phytoplankton cells, is the research area that has advanced rapidly in the past decade, and is reviewed extensively. Key steps of quantitative immunofluorescence, fixation and immunolabel intensity, are discussed in more detail. Qualitative immunofluorescence is a new, hardly explored but highly interesting development in which qualitative physiological variables related to e.g. nutrient limitation or primary production are measured in individual cells instead of phytoplankton populations as a whole. Several combinations of immunological probes, both for species identification and for physiological measurements, are proposed. A special case of qualitative immunofluorescence is the measurement of phytoplankton toxins in single cells from natural populations. It is anticipated that the future use of semiconductor nanocrystals or “quantum dots” as fluorophores will greatly enhance signal detection in flow cytometry, and hence in both quantitative and qualitative immunofluorescence applications.

**Key words:** flow cytometry, immunology, marine phytoplankton, fluorescence, European Optical Plankton Analyser (EurOPA), physiology, antibodies

**INTRODUCTION**

Immunology and flow cytometry are becoming powerful tools in the study of phytoplankton ecology. The application of immunology in phytoplankton research started thirty years ago when Bernhard et al. (1969) developed antibodies against the diatoms *Leptocylindrus danicus* and *Phaeodactylum tricornutum*. It was not, however, until the 1980’s that immunological techniques for species identification were actually applied in marine research. The first species investigated were prokaryotes (Dahl and Laake, 1982; Campbell et al., 1983), but some years later eukaryotic organisms followed (Hiroish et al., 1988; Shapiro et al., 1989). The swift development in the past decade of species-specific labelling with antibodies, especially for the identification of harmful algae, has been reviewed by Anderson (1995). A more general review on immunofluorescence in phytoplankton research was presented by Vrieling and Anderson (1996).

During the development of immunological techniques, the flow cytometer, a device that counts and measures cells in single file, became widely applied in aquatic research (Yentsch and Horan, 1989). The combination of flow cytometry and immunofluores-
cence as a new tool to identify and count target cells in natural seston assemblages was outlined by Cucci and Robbins (1988). However, at that time this technique had not been applied in aquatic science yet.

In this paper the development of immunological techniques in marine eukaryotic phytoplankton research since Cucci and Robbins (1988) will be reviewed, focusing on the combination of immunology and flow cytometry. Firstly, the basics of immunochemical methods and flow cytometry will be explained. Then, the species-specific immunological probing, and quantitative immunofluorescence flow cytometry will be summarised. Finally, the possibilities of antibody application in physiological research will be described.

**IMMUNISATION, ANTIBODIES AND IMMUNOCHEMICAL METHODS**

Three important issues are distinguished for the immunolabelling of phytoplankton cells. Firstly, antibodies for the identification of specific molecules or cell types are essential, but they can only be obtained when the antigen is defined. Secondly, a selective reporter molecule, a fluorescent probe or fluorophore, should be selected for the adequate detection of antigens or cells in samples with multiple fluorescence signals. Finally, an efficient procedure is required for reproducible labelling at a most optimal intensity of the fluorescent signal of the reporter molecule.

**Immunisation**

The preparation of antibodies requires a proper definition of the antigen and the most useful antibodies are obtained using highly purified antigens. Good examples are antibodies that recognise specific cellular molecules of phytoplankton cells, which are of physiological or metabolic importance such as the carbon-fixing enzyme Rubisco (Orellana and Perry, 1992, 1995). Sometimes, however, the purification of the antigen is not possible or even not required as in the case of antibodies to cell surface components. In this case the whole cell, or cell wall fragments in a purified fraction, can be considered the antigen. In phytoplankton research whole cells or cell wall fragments, sometimes stabilised by aldehyde fixation, have often been used for small species such as *Synechococcus* and *Aureococcus* (Campbell et al., 1983; Anderson et al., 1989) as well as larger species such as dinoflagellates of the genera *Alexandrium*, *Gymnodinium*, and *Prorocentrum*, diatoms of the genus *Pseudo-nitzschia*, and raphidophytes of the genus *Chattonella* (Vrieling and Anderson, 1996 and references therein). Table 1 provides a detailed list of all eukaryotic marine phytoplankton species for which antibodies have been produced in the past decade.

After design of the antigen a host animal should be selected. Most commonly, rabbits are used but depending on the amount of antibodies needed, one can decide to use larger animals. For the establishment of “immortal” antibody-producing cell lines mice are immunised. Immunisation proceeds by repetitive injections of the antigen in the host animals. In general, antibodies are obtained within 6 months.

**Antibodies**

Antibodies are immunoglobulins that are secreted from activated lymphocytes when host organisms are exposed to foreign molecules or micro-organisms. Immunoglobulins are large molecules (150 to 950 kDa) with a common feature of protein configurations; each containing identical short and large peptide chains that are known as light and heavy chains. Two heavy and light chains together form a typical Y-shaped molecule (Fig. 1). The molecule is divided into domains of which the regions where targets bind are highly variable. Adjacent variable domains of both the heavy and the light chain form the antigen binding sites (the Fab regions) per antibody molecule. The other domains, of which the tail end is also known as the Fe-region, are constant or conserved. The type of heavy chain distinguishes immunoglobulins in the classes IgA, IgG, IgD, IgE, and IgM, of which IgA and IgM are di- and pentamers, respectively. In principle each lymphocyte produces antibodies that recognise one epitope of the antigen. The total serum of an immunised animal will contain many different antibodies that consequently is called a polyclonal antiserum (often referred to as polyclonal antibodies or PAbs). Uni-epitope specific or monoclonal antibodies (MAbs) can be obtained by fusing spleen cells of an immunised animal with myeloma cells, from which multiple hybridoma cell lines are established, each of which produces antibodies against one single epitope. A disadvantage of MAb’s is that, in case of fixation, this single epitope may be destroyed so that a MAb will fail to bind, while a PAb would still recognise other epitopes.
Table 1. – Marine eukaryotic phytoplankton species identification with immunological techniques. In cases where flow cytometry (fcm) was not applied (no), epifluorescence microscopy or magnetic beads were used.

| Class                          | species                              | culture/field application | PAb/ MAb | fcm applied | reference                  |
|-------------------------------|--------------------------------------|---------------------------|----------|-------------|----------------------------|
| Bacillariophyceae             | *Pseudo-nitzschia multiseries*       | culture/Wadden Sea        | PAb<sup>trc</sup> | no          | Bates et al., 1993, Vrieling et al., 1996b |
|                               | *Pseudo-nitzschia multiseries*       | culture                   | MAb<sup>APC</sup> | EurOPA      | L. Peperzak & B. Sandee, unpublished |
|                               | *Pseudo-nitzschia pungens*           | culture/Wadden Sea        | PAb<sup>trc</sup> | no          | Bates et al., 1993, Vrieling et al., 1996b |
|                               | *Thalassiosira oceanica*              | culture/Atlantic Ocean    | PAb<sup>trc</sup> | no          | Shapiro et al., 1989, Campbell et al., 1994 |
| Chlorophyceae                 | *Dunaliea tertiolecta*                | culture                   | PAb<sup>trc</sup> | no          | Shapiro et al., 1989 |
|                               | B6125 (unidentified)                 | Atlantic Ocean            | PAb<sup>trc</sup> | no          | Campbell et al., 1994 |
| Cryptophyceae                 | *Chroomonas salina*                  | culture                   | PAb<sup>trc</sup> | no          | Shapiro et al., 1989 |
| Dinophyceae                   | *Alexandrium fundyense*               | culture/Gulf of Maine     | MAb<sup>APC</sup> | EPICS ELITE<sup>a,b</sup> | Sako et al., 1993, Sako et al., 1996 |
|                               | *Alexandrium lusitanicum*             | culture/Spainish coast    | PAb<sup>trc</sup> | no          | Mendoza et al., 1995 |
|                               | *Alexandrium minutum*                 | culture/Spainish coast    | PAb<sup>trc</sup> | no          | Mendoza et al., 1995, Costas and Lopez-Rodas 1996 |
|                               | *Alexandrium tamarense*               | culture/New Zealand coast | PAb<sup>trc</sup> | no          | Chang et al., 1999 |
|                               | *Alexandrium tamarense*               | culture                   | MAb<sup>APC</sup> | FACS STAR<sup>b</sup> | Taylor and Lewis 1995 |
|                               | *Alexandrium tamarense*               | culture/japanese coast    | MAb<sup>APC</sup> | no          | Adachi et al., 1993 |
|                               | *Alexandrium tamarense*               | North Sea                 | MAb<sup>APC</sup> | no          | Sako et al., 1993, Sako et al., 1996 |
|                               | *Gymnodinium catenatum*               | culture/Spainish coast    | PAb<sup>trc</sup> | no          | Mendoza et al., 1995 |
|                               | G. mikimotoi<sup>a</sup>             | culture/central North Sea | PAb<sup>trc</sup> | no          | Mendoza et al., 1993 |
|                               | G. mikimotoi<sup>b</sup>             | North Sea                 | MAb<sup>APC</sup> | no/EurOPA   | L. Peperzak & B. Sandee, unpubl. |
|                               | *G. nagasakiense*                     | culture                   | MAb<sup>APC</sup> | no          | Nagasaki et al., 1991 |
|                               | *Gymnodinium simplex*                 | culture                   | PAb<sup>APC</sup> | OPA         | Vrieling et al., 1994, 1995a, 1996a |
|                               | *Prymnesium parvulum*                 | culture/Atlantic Ocean    | PAb<sup>APC</sup> | no          | Vrieling et al., 1994, 1995a, 1996a |
|                               | *Prorocentrum lima*                   | culture/Spainish coast    | PAb<sup>APC</sup> | no          | Vrieling et al., 1994, 1995a, 1996a |
|                               | *Prorocentrum micans*                 | culture                   | PAb<sup>APC</sup> | no          | Vrieling et al., 1994, 1995a, 1996a |
|                               | *Prorocentrum minimum*                | culture                   | PAb<sup>APC</sup> | no          | Vrieling et al., 1994, 1995a, 1996a |
|                               | *Prorocentrum sp*                     | culture                   | PAb<sup>APC</sup> | no          | Vrieling et al., 1994, 1995a, 1996a |
|                               | *Gyrodinium aureolum*                 | USA NE coast              | PAb<sup>APC</sup> | no          | Anderson et al., 1989, 1993 |
|                               | *Aureoumbra lagunensis*               | USA SE coast              | PAb<sup>APC</sup> | no          | Lopez-Barreiro et al., 1998 |
|                               | *Pseudonitzschia multiseries*         | culture/Atlantic Ocean    | PAb<sup>APC</sup> | no          | Shapiro et al., 1989, Campbell et al., 1994 |
|                               | *Aureococcus anophagefferens*         | USA NE coast              | PAb<sup>APC</sup> | no          | Anderson et al., 1989, 1993 |
|                               | *Pseudonitzschia pungens*             | culture/Atlantic Ocean    | PAb<sup>APC</sup> | no          | Lopez-Barreiro et al., 1998 |
|                               | *Thalassiosira oceanica*              | culture/Atlantic Ocean    | PAb<sup>APC</sup> | no          | Shapiro et al., 1989, Campbell et al., 1994 |
|                               | *Plymnesion parvulum*                 | culture/Atlantic Ocean    | PAb<sup>APC</sup> | no          | Shapiro et al., 1989, Campbell et al., 1994 |
|                               | *Emiliania huxleyi*                   | North Sea/Atlantic Ocean  | PAb<sup>APC</sup> | no          | van Bleijswijk et al., 1991 |
|                               | *Chattonella antiqua*                 | culture/ North and Wadden Sea | MAb<sup>APC</sup> | no          | Hiroshi et al., 1988, Vrieling et al., 1995b |
|                               | *Chattonella marina*                  | culture                   | MAb<sup>APC</sup> | no          | Hiroshi et al., 1988 |

<sup>a</sup>Immunologically identical to *Gyrodinium aureolum* and *Gymnodinium nagasakiense*
<sup>b</sup>Unsuccessfully applied in field samples
<sup>c</sup>PAb / MAb = Polyclonal / Monoclonal Antibody with fluorophore in superscript; FT<sup>c</sup>c = conjugated using FluoresceinIsotioCyanate; AP<sup>c</sup>c = AlloPhycocyanin
<sup>d</sup>EPICS<sup>c</sup>c = Coulter, FACS<sup>c</sup>c = Becton Dickinson, OPA = Optical Plankton Analyser, EurOPA = European Optical Plankton Analyser
<sup>e</sup>Texas Brown Tide organism
<sup>f</sup>MAb GN-89 did not label a Portuguese strain of *G. cf. nagasakiense* (*Gyrodinium cf. aureolum*)
<sup>g</sup>Several strain-specific MAb’s

Other differences between polyclonal and monoclonal antibodies are the financial costs and the amount of antibodies that can be obtained. The easiest way is the preparation of PABs, since it is quite easy, quick (immunisation of about 3 months), and relatively cheap (< $500). The disadvantage is the relatively low amount of antibodies retrieved from a short-lived animal, which requires immunisation of new animals and repeated checking of antibody specificity and validity. The advantage of establishing antibody-producing cell lines or MAbs is that by careful screening clones can be selected that each produce antibodies recognising a highly specific epitope. Furthermore, these cell lines continuously produce anti-
bodies in tissue cultures, and they can be stored in liquid nitrogen allowing production of new and identical batches whenever desired: the so-called “immortal” cell lines.

After obtaining antibodies, they should be tested for reactivity and specificity at various levels. The first step is determining the antigen binding capacity in a dilution series to determine the optimal concentration of the primary antibodies for use, or vice versa the required concentration of antigens at a given antibody concentration. The immune-response may vary by the pre-treatment of the samples.

**Immunological methods**

Complementary to the prepared antibodies are the desired reporter molecules for identification. For immuno flow cytometry only fluorescent probes are applicable. The antigens can be identified using either direct or indirect labelling, while the signal/noise-ratio of the indirect method can be improved by labelling enhancement procedures (Vrieling et al., 1993b). Direct immuno-labelling (Fig. 1) implies that the fluorescent molecule is directly attached to the antibody that recognises the antigen (the primary antibody). This method not only results in a lower background, but can be performed using fewer reagents and washing steps which are important considerations in quantification studies (Vrieling et al., 1996a, 1997). Direct immunofluorescence is easy for labelling the cell surface of phytoplankton cells. However, the labelling of intracellular antigens may not always be possible due to the charge or size of the antibodies. In most phytoplankton studies, the probe for identification has been linked to another antibody (a secondary antibody), that specifically binds to the applied primary antibody. If for example the primary antibody is raised in rabbits, the secondary should be an anti-rabbit antibody. These secondary antibodies can be purchased from various companies and multiple types are available with respect to the reactivity to the host of the primary antibodies, or the desired fluorescent reporter molecules. It is surprising, however, that from all available fluorescent molecules, the use of fluorescein isothyocyanate (FITC) dominates in immunofluorescence studies of phytoplankton (Table 1).

When the label intensity is not sufficient for the proper detection of a cell, the signal can be enhanced. In that case the primary or secondary antibody most often is covalently linked to several molecules of the vitamin biotin. By the use of fluorescently labelled (strept) avidin, compounds that have a very strong affinity for biotin, a complex formation of fluorescent molecules is realised (Fig. 1).

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**Fig. 1.** Methods of immunofluorescence: direct, indirect and enhanced. Antibodies are Y-shaped: black = primary, white = secondary antibody. Circles with stars are fluorescent labels or fluorophores. In the direct assay the fluorophore is bound to the primary antibody. In the indirect assay, the fluorophore is bound to the secondary antibody. In the enhanced assay, several molecules of biotin are linked to the secondary antibody. To the biotin fluorescent streptavidin is bound; this last step can be used to make fluorescent complexes. One streptavidin molecule can be linked directly with several fluorophores.
These complexes result in an increase of the fluorescence intensity over common direct or indirect labelling procedures (Vrieling et al., 1993b).

Compared to direct immunofluorescence, the indirect methods for intracellular labelling are not importantly affected by permeability problems. Since multiple fluorescent molecules are coupled to the antigen the labelling intensity often increases. Furthermore, indirect assays are flexible with respect to the fluorescent reporter molecule used. Some disadvantages should also be considered. The number of wash and incubation steps increases in indirect assays, resulting in extra time and cell loss. Furthermore, additional background labelling as a result of non-specific binding of secondary antibodies may be introduced and more control measurements (“negative controls”) should be performed.

FLOW CYTOMETRY AND THE EUROPEAN OPTICAL PLANKTON ANALYSER

Flow cytometry is applied for the determination of phytoplankton abundance for several reasons. Firstly, microscopic phytoplankton enumeration is labour intensive and is often not objective, cells smaller than about 4 µm are not easy to discriminate and to enumerate. Secondly, the biomass of cell populations is difficult to assess accurately, because biomass is usually calculated by multiplication of the cell number and an overall estimate of the species cell volume. Thirdly, the amount of microscopic data obtained, e.g. in monitoring programmes is large and detailed and before these data can be interpreted, they usually are clustered into the main taxonomic groups (e.g. diatoms, flagellates). In contrast, flow cytometry allows the rapid enumeration of (small) cells, size measurements of large numbers of cells, and the combination of size with autofluorescent properties of individually detected cells can be used in cluster analyses.

Most oceanographic applications like the counting of small sized oceanic phytoplankton allow the use of standard commercial flow cytometers. However, coastal and inland species are not only highly heterogeneous in taxonomic composition, size, concentration, and morphology, but many also form colonies. Therefore, an optical plankton analyser (OPA; Peeters et al., 1989; Dubelaar et al., 1989) and later a more compact version, the European optical plankton analyser (EurOPA; Jonker et al., 1995) was designed and built.

The principle of flow cytometry, and the EurOPA in particular, is shortly described. A suspension of particles or (un)treated phytoplankton cells is sampled and is brought in a sheath flow resulting in a hydrodynamically focused sample stream (Fig. 2). The particles in this stream are transferred as a sin-

![Fig. 2. – Schematic drawing of a dual beam flow cytometer. The sample, containing 3 different kinds of particles (1. black = non-fluorescent, 2. white + nucleus = not species-specifically labelled but with a stained nucleus, 3. white + nucleus + 4 fluorophores) is pumped into the particle-free sheath-fluid. The sample stream is contained in the core (“hydrodynamic focusing”). The particles pass 2 laser beams in single file. Under a 90° angle the fluorescent light is detected by “photondetectors”, e.g. photomultipliers after the scattered laser light has been filtered. Unfiltered 90° scattered laser light is called perpendicular light scatter (PLS) and is a measure of particle morphology and structure. Forward scattered laser light (FLS) is a measure of particle size. In the case of measuring the nuclear DNA concentration of G. mikimotoi cells (see Fig. 3), laser 1: $\lambda = 529$ nm, laser 2: $\lambda = 633$ nm, filters 1 and 2: $\lambda > 670$ nm. The signals from the photondetectors are processed with electronic hardware and a computer to produce so-called dot-plots.](image-url)
gle particle file with a constant velocity to one or two focused laser beams. A list of commonly used lasers is provided in Table 2.

Laser light is scattered when a particle crosses the laser beam. Forward scattered light (FLS) under very narrow forward angles correlates with particle size and shape according to Mie’s derived Fraunhofer theory for small particles in the range of 2-20 µm (Cunningham and Buonaccorsi, 1992). Perpendicular light scatter (PLS) represents particle morphology and structure (Dubelaar et al., 1987; Ackelson et al., 1988). Laser light that irradiates chlorophyll containing phytoplankton is absorbed by the antenna pigments, and the excitation energy is transferred to the reaction centres of their two photosystems, that emit fluorescence (Krause and Weis, 1991). An additional process is the emission of fluorescence by accessory pigments (Yentsch and Yentsch, 1979; Yentsch and Phinney, 1988; Hofstraat et al., 1991). Fluorescence emission excited by a green or blue laser is measured in perpendicular direction in the red region (FGR (G=Green laser) or FBO (B=Blue laser): chlorophyll a) and in the orange bandwidth (FGO or FBR: phycocerythrin, phycocyanin). In the EurOPA the fluorescence emission of autotrophic phytoplankton cells excited by the second, red laser is measured (perpendicularly) in the red region (FRR, chlorophyll-b, phycocyanin). Typical excitation and emission wavelengths in flow cytometry are listed in Table 2.

The laser beam passage time, or the time of flight (TOF), corresponds with the particle length and is measured in the longitudinal direction, ranging from 1-1000 µm by the EurOPA. Commercial flow cytometers can only measure cells up to 50 µm. The size of the scatter and fluorescence signals are integrated, and the measurements are stored in a data list ranging from 1-10^8 arbitrary units.

Unfortunately, due to small differences in size and in pigment composition between many phytoplankton species, the flow cytometric discrimination in multi-species field samples is, unlike microscopy, still difficult (Hofstraat et al., 1994). In order to identify one species of interest in such samples, e.g. a harmful algal species, a program was started to produce species-specific antibodies. The antibodies

**Table 2.** Excitation (EX) and emission (EM) wavelengths of the Optical Plankton Analyser (OPA), the European OPA (EurOPA) and commercially available flow cytometers, as listed in Table 1. PLS = Perpendicular Light Scatter, FLS = Forward Light Scatter, F = Fluorescence, B = Blue, G = Green, R = Red. In dual beam flow cytometers such as the OPA and EurOPA a particle can be investigated with 2 lasers, e.g. green and red (see Fig. 2).

| Flow cytometer | Excitation lasers (λ_ex) | FCM variable | Emission (λ_em) |
|----------------|--------------------------|--------------|----------------|
| OPA, EurOPA    | Blue: 488 nm (Ar-gas)^b  | PLSB         | < 495 nm       |
|                |                          | FLSB         | < 495 nm       |
|                |                          | FBG          | 500-550 nm     |
| EPICS CS       |                          | FBO          | 550-630 nm     |
|                |                          | FBR          | ≥ 670 nm       |
| FACS STAR      | Green: 529 nm (Ar-gas)^b | PLSG         | < 535 nm       |
|                | or: Green: 532 nm (NdYAG)^c | FLSG     | < 535 nm       |
|                |                          | FGO          | 550-630 nm     |
|                |                          | FGR          | ≥ 670 nm       |
| OPA, EurOPA    | Red: 633 nm (HeNe)^d    | FRR          | ≥ 670 nm       |

^a Eur(OPA) emission wavelengths
^b Argon, gas laser
^c Neodymium Yttrium Aluminium Garnet, solid state laser
^d Helium Neon, gas laser

**Table 3.** Excitation and emission wavelengths in nanometers of the fluorophores mentioned in the text, and of phytoplankton autofluorescence. Data from Shapiro 1985 and Vrieling and Anderson (1996).

| fluorophore         | abbreviation | Excitation range | Excitation maximum | Emission range | Emission maximum |
|---------------------|--------------|------------------|--------------------|----------------|------------------|
| Chlorophyll-a       | Chla         | 340-630          | 440                | 650-750        | 670              |
| Calcoflour white    | Calcofluor   | 340-360          | -                  | 400-440        | -                |
| Autofluorescent cell wall | -        | -                | 460                | -              | 535              |
| FluorescinIsoThioCyanate | FITC      | 420-520          | 490                | 490-590        | 520              |
| PropidiumIodide     | PI           | 450-690          | 540                | 550-710        | 610              |
| PhycoErythrin       | PE           | 460-590          | 550                | 550-630        | 580              |
| Allo PhycoCyanin    | APC          | 520-670          | 650                | 640-720        | 670              |

170 L. PEPERZAK et al.
were labelled with fluorophores to enable the identification and enumeration using flow cytometry. Table 3 provides excitation and emission specifications of commonly used fluorophores. Besides fluorescent antibodies, specific cell components, such as DNA, can also be labelled with probes that penetrate living or fixed cells (Table 3).

IMMUNOFLUORESCENT SPECIES IDENTIFICATION

In recent years the number of antibodies against marine eukaryotic phytoplankton species has increased considerably. To date there are antibodies against 28 species in 8 phytoplankton classes (Table 1). Some of the antibodies are directed against species which are microscopically difficult to identify such as small pelagophytes and prasinophytes. Several dinoflagellate species are able to produce potent neurotoxins (Hallegraeff, 1993) and this feature has intensified research in this class: as a result antibodies, 26 are polyclonal and only 10 are monoclonal antibodies have been developed against both species (Hiroishi et al., 1988). The genus Pseudo-nitzschia once contained the non-toxic P. pungens f. pungens and the toxic P. pungens f. multiseries. The two forms could not be discriminated with the light microscope and only with electron microscopy accurate taxonomic classification was possible. At present the two Pseudo-nitzschia forms are considered two species (Tomas, 1996), and they can be differentiated with the fluorescence microscope using polyclonal antibodies (Bates et al., 1993).

Most widely used in immunological species identification are polyclonal antibodies. Of a total of 36 antibodies, 26 are polyclonal and only 10 are monoclonal (Table 1). Two polyclonals have been made against Alexandrium minutum, Emiliania huxleyi and Prorocentrum micans. Against Alexandrium tamarense, one polyclonal and several monoclonal antibodies, and for Gymnodinium mikimotoi, considered to be conspecific to Gyrodinium aureolum and Gymnodinium nagasakiense (Gentien, 1998), two monoclonal antibodies have been developed.

Newly produced antibodies are first checked using laboratory cultures before using them in field samples. Some of the antibodies, however, have not yet been applied to field samples (Table 1). Whether applied in cultures or in field samples, the majority of research has been carried out using epifluorescence microscopy. Flow cytometry has been performed with commercial flow cytometers (Coulter Counter, Becton Dickinson) and with dedicated flow cytometers such as the OPA and EurOPA (Table 1). The application of flow cytometry in combination with immunofluorescence will be described in more detail in the next section.

QUANTITATIVE IMMUNO FLOW CYTOMETRY

One of the prime objectives of immunochemical labelling is the fast counting of target cells in field samples. The most widely applied method to concentrate, to label and to visualise phytoplankton cells is using membrane filters and epifluorescence microscopy (Table 1). One of the advantages of membrane filtration is that the chance that cells will be lost in the various incubation steps necessary for immunochemical labelling is limited. On the other hand, immunochemical labelling for flow cytometry is performed in tubes in which the cells are repeatedly centrifuged and resuspended before they can be analysed. Furthermore, a flow cytometer is more expensive than an epifluorescence microscope and this is the reason why the combination of immunolabelling and flow cytometry has been tested in a few studies only (Table 1). However, when immuno flow cytometry is feasible, counting can be fast and flow cytometric applications, e.g. size measurements and cell sorting, can be performed with the same speed. Here, examples and quantitative aspects of immuno flow cytometry will be presented.

To our knowledge, the first flow cytometric analysis of immunolabelled eukaryotic marine phytoplankton species was performed on P. micans and G. mikimotoi with the Optical Plankton Analyzer or OPA (Vrieling et al., 1993a). A species-specific and a non-specific polyclonal rabbit IgG antibody lead to small (P. micans) and somewhat larger (G. mikimotoi) increases in green fluorescence using a FITC-conjugated goat-anti-rabbit IgG antibody (GAR-FITC). Subsequently, labelling intensity was enhanced using a biotin-streptavidin system (Vrieling et al., 1993b; Fig. 1). To the primary polyclonal rabbit antibody a secondary donkey-anti-rabbit IgG-biotin (DAR-biotin) was bound. The biotinylated secondary antibody was bound with streptavidin-FITC, leading to a doubling of green fluorescence intensity as measured with the OPA in comparison with the indirect method (Vrieling et al., 1993b).
The OPA flow cytometer was also used in measurements of *G. mikimotoi* testing several monoclonal antibodies prepared against this species (Vrieling *et al*., 1994). Such measurements were performed in samples of cultures, artificially mixed algal populations as well as in field samples (Vrieling *et al*., 1994, 1995a, 1996a). In two year old, Lugol-fixed field samples *G. mikimotoi* could still be counted at concentrations of less than 2000 cells litre\(^{-1}\) using a direct monoclonal-FITC label. Although such a direct labelling reduces the number of wash steps in the assay, the recovery of *G. mikimotoi* cells was low (0-20\%) as compared to light microscopic counts (Vrieling *et al*., 1995a). On the other hand, when properly, paraformaldehyde-fixed *G. mikimotoi* cells in mixed algal populations were labelled directly, and measured with the OPA gated on FBG and FBO, the recovery increased to 77\% (Vrieling *et al*., 1996a).

A special phytoplankton flow cytometer like the OPA is not a prerequisite for measuring potentially toxic, single-celled dinoflagellates, which all are smaller than 100 \(\mu\)m. Therefore, commercially available flow cytometers have been used successfully too in immunochemical labelling experiments. The thecate genus *Alexandrium* has received considerable attention, probably because of its potentially toxic effects on a wide variety of animals, including humans (Hallegraeff, 1993). Taylor and Lewis (1995) measured *A. tamarense* cells labelled with a monoclonal antibody probed with a FITC-conjugated donkey anti-rabbit antibody. Sako *et al.* (1996) produced monoclonal antibodies against Japanese isolates of *Alexandrium catenella* and *A. tamarense*. Labelling took place with goat anti-mouse IgG-FITC or indirect with biotinylated goat anti-mouse IgG and streptavidin-FITC in order to enhance label intensity. Fixation of cultures with 2\% formaldehyde produced the strongest fluorescence intensities, and the highest cell recovery (70\%) in flow cytometric enumeration compared to direct microscope counts. This study by Sako *et al.* (1996) is the first attempt to measure *Alexandrium* spp. in field samples using flow cytometry. However, the quantification of *Alexandrium* spp. was problematic, probably as a results of cell loss during plankton collection and interference from electric noise during analysis (Sako *et al.*, 1996).

A special combination of immunological labelling, using either a fluorophore or magnetic beads, was performed by Costas and Lopez-Rodas (1996). The polyclonal antibody had been produced against *Alexandrium minutum* but it was genus (*Alexandrium*) specific (Mendoza *et al.*, 1995). The antibody was first made *A. minutum* specific by pre-incubating it with *Alexandrium excavatum* and *Alexandrium affine* cells. Field samples (fixed with 4\% buffered formaldehyde) were then treated with the pre-incubated polyclonal in an indirect assay. The secondary antibody was either FITC-conjugated goat anti-rabbit or magnetic beads-conjugated goat anti-rabbit antibody (Costas and Lopez-Rodas, 1996). The FITC-labelled cells were counted with a flow cytometer while bead-labelled cells were isolated using a magnet. The recovery of the FITC-labelled *A. minutum* cells in undiluted seawater samples was quite high (75-99\%), while the bead-labelled cells were also efficiently separated with recoveries ranging from 61-92\% (Costas and Lopez-Rodas, 1996).

**Difficulties in quantitative immunological flow cytometry**

There are two major problems in quantifying immunologically labelled cells in flow cytometry. Firstly, the fixation and labelling procedure involves numerous steps (incubation, washing, centrifugation) that can lead to cell loss. Secondly, the label intensity should be at such a level that the target cells produce a signal that is specific and strong enough for them to be discriminated against non-target cells. Since these problems co-occur in immuno flow cytometry, they need serious attention; they will be discussed in more detail.

**Fixation and labelling procedure**

Antibodies are produced in animals or in immortal cell lines at a salinity typically 5 times lower than seawater. The phytoplankton cells have to be transferred to this lower salinity before antibodies can be applied. In order to prevent cell disruption due to osmosis the cells have to be fixed, but ironically fixation also leads to some cell loss. The fixation problem is not new, nor can it be avoided. It is necessary, therefore, to quantify and minimise the artefacts of fixation, and to compare flow cytometric counts with conventional microscopic counts. But first, an appropriate fixation method must be chosen.

One of the most widely used fixatives in quantitative phytoplankton studies is Lugol, an acid iodine solution. Although the monoclonal anti-*G. mikimotoi* antibody GA8 still recognised Lugol-fixed cells,
the recovery of the immunological counts were low (Vrieling et al. 1995a) and, therefore, Lugol is not preferable in quantitative studies.

Aldehyde fixatives cross-link proteins, they are colourless and they are routinely used in immunology. Glutaraldehyde is commonly used as a fixative in electron microscopy, but it increases green autofluorescence of phytoplankton cells (Shapiro et al., 1989) and it can reduce antigen binding (Vrieling et al., 1994). Furthermore, glutaraldehyde may lead to false positives, because it can bind IgG to its free active sites (Vrieling and Anderson, 1996). Formaldehyde or ‘paraformaldehyde’ (= formaldehyde made from freshly depolymerised paraformaldehyde) at 2-4 % (w/v) concentrations usually lead to good antigen binding capacity. The reason probably is, that the phytoplankton cells that are used to immunise the host animal usually are fixed in the same way. Flow cytometric cell recoveries with formaldehyde fixatives range from 70-99% (Costas and Lopez-Rodas, 1996; Sako et al., 1996, Vrieling et al., 1996a). Low concentrations (0.01-0.1%) of paraformaldehyde may be sufficient for intracellular antibody assays (Navaluna et al., 1989); additional permeabilisation of the cell wall using ethanol (Orellana and Perry, 1995) may be necessary. In practice the best fixative, and its optimum concentration, have to be tested experimentally for each species (e.g. Lin and Carpenter, 1996a).

The number of incubation, washing and centrifugation steps will depend on the type of immuno-assay employed (Fig. 1). The optimum protocol, which minimises cell loss and maximises label intensity, must be determined experimentally as well. For instance, the necessary antibody concentration can be measured in a dilution series of the target cells (Orellana and Perry, 1995; Vrieling et al., 1997).

Label intensity

The emission wavelength of the fluorophore and its intensity, should be such that the target cells are readily identified and counted. In flow cytometry this means, that the label should be excitable with commonly used lasers (e.g. the Argon laser, Table 2), that quantum yield should be high, and the emitted light should have a narrow emission spectrum. In both immuno flow cytometry and epifluorescence microscopy, FluoresceinIsoThioCyanate (FITC) is used almost exclusively as fluorophore (Table 1). Much immunochemical research has been carried out using FITC as a label, and FITC can be excited over a wide range of wavelengths including the 488 nm line of argon lasers (Table 2, 3). The problem of FITC is that it emits at the same wavelengths as the auto-fluorescent phytoplankton cell wall (Table 3). Phycoerythrin has an excitation maximum at 460 nm and can also be excited with an Argon laser (Table 3). Phycoerythrin has been used in epifluorescence dual (FITC and PE) staining techniques (Lin and Carpenter, 1996a) but not in flow cytometric applications. AlloPhycoCyanin (APC) is a red fluorescent fluorophore (Table 3) that needs exciting at 630 nm which means the need for a HeNe laser (Table 2), and bleaching of the chlorophyll’s. APC has been used in indirect immunofluorescent techniques, in which P. multiseries was counted flow cytometrically (Table I) or in which G. mikimotoi DNA was simultaneously stained with propidium iodide (Fig. 3).
A new type of secondary antibody labels are the so-called Alexa™-conjugates. According to the manufacturer, Molecular Probes, these conjugates produce a brighter fluorescence than conventional fluorophores. Additionally, so-called tandem-dyes (Alexa 488 - APC) are available. When such a tandem-dye is excited at 488 nm, it emits at 660 nm. This means that dual-labelled cells (e.g. with FITC and Alexa 488-APC) can be excited with only one laser. However, at present these new labels have not been tested in immuno flow cytometry.

Quantum dots, or semiconductor nanocrystals offer a new class of fluorophores. These are nanometer-sized semi-conductor particles with a broad excitation range and a narrow emission range that is dependent on the particle size (Chan and Nie, 1998; Bruchez et al., 1998). One wavelength can therefore be used to excite many differently sized quantum dots, yielding many emission wavelengths simultaneously. Furthermore, the emission spectra are very narrow, the emission intensity is very high and, important in epifluorescence microscopy, the quantum dots appear to be much more stable against photobleaching (Chan and Nie, 1998; Bruchez et al., 1998). In this case, however, quantum dots are so new that they are not yet available commercially.

QUALITATIVE IMMUNO FLOW CYTOMETRY

In the previous section the emphasis lay on quantitative species recognition and quantification. The counting of phytoplankton species in a sample of water is one of the first steps in the study of aquatic ecosystems. The measurement of qualitative aspects, such as size and composition of individual cells or populations, is a logical second step. Such measurements can, in principle, be made with a flow cytometer. Measurements of size can be made with any flow cytometer, e.g. as scatter or as time of flight. The composition of the cells can be analysed flow cytometrically only when the desired components are specifically stained. With regard to immunofluorescence, three combinations are possible:

(i) immunofluorescent species identification + non-immunological fluorophore,
(ii) immunofluorescent species identification + immunofluorescent fluorophore,
(iii) non-immunological species identification + immunofluorescent fluorophore.

The fourth combination, in which both fluorophores are non-immunological, are outside the scope of this paper. Of each of the three above mentioned combinations examples will be provided. Unfortunately, the number of practical applications is still limited. Therefore, some of the applications are still theoretical, but they are given to illustrate the potential of flow cytometry in combination with immunofluorescence. As with quantitative flow cytometry, good care has to be taken in cell fixation. When cell components are labelled with an immunological fluorophore, the cells need to be permeabilised so that the antibody can pass the plasmalemma. With respect to this problem Orellana and Perry (1995) and Lin and Carpenter (1996a) found 96% ethanol to be a good fixative. However, the effect of this method of fixation on cell surface, species-specific, antigens has not been tested yet.

Immunofluorescent species identification + non-immunological fluorophore

When it is possible to identify a phytoplankton species in a mixed assemblage, additional measurements on its cell-composition can provide worthwhile ecological information. The simplest example is the simultaneous measurement of the cells autofluorescence. The red fluorescence intensity is related to the cellular chlorophyll-a content. For instance, the red chlorophyll-a autofluorescence of immunochemically (FITC)-labelled G. mikimotoi cells has been used in cluster analysis (Vrieling et al., 1994). In living phytoplankton, chlorophyll-a fluorescence is not a simple measure of chlorophyll-a, because the fluorescence yield is not constant but depends on the cells irradiance history and physiological status (Kiefer, 1973; Slovacek and Hanman, 1977). For immunological purposes, the cells will be fixed and that may stabilise the fluorescence yield. Unfortunately, Navaluna et al. (1989) found that the red chlorophyll fluorescence of 0.01% paraformaldehyde fixed cells declines rapidly and significantly in the first 100 hours after fixation. At the moment it is unclear how the immunofluorescence protocol, which also includes several labelling and wash steps, will impair reliable chlorophyll-a measurements.

A second example of immunofluorescent labelling in combination with a non-immunological fluorophore, stems from the flow cytometric measurement of in situ growth rates. Phytoplankton growth rates in the sea can be calculated from temporal, daily or weekly, changes in cell concentra-
tions. The problems in this approach are that net changes and net growth rates (not intrinsic rates) are calculated, and that the body of water at the sampling location will change in time. The in situ production rate measurements where samples are incubated in bottles, can lead to errors due to the so-called “bottle enclosure effect” (Gieskes et al., 1979). These problems are overcome, when the population of a given species is sampled repeatedly in the same body of water, and if the proportion of the cells that undergo phased cell division in a diel cycle can be measured. Before cell division can take place the DNA in the nucleus must have doubled and, at least in dinoflagellates, this process can be quantified relatively easy. This is the basis of the in situ diel DNA synthesis growth rate determination as described and tested by Carpenter and Chang (1988) and Chang and Carpenter (1988). Because this growth rate determination is based on the proportion of the cells that divide, the method does not require exact quantification of cells. However, a major problem in the original method was the large number of manual microscope DNA measurements that had to be made: 320 cells in 13 samples (Chang and Carpenter, 1988). Flow cytometric DNA measurements, after SYTOX, Pico Green or propidium iodide staining, are made much more rapidly (Veldhuis et al., 1997), and the results can readily be applied to the Carpenter-Chang model (Pan and Cembella, 1998; Peperzak et al., 1998b).

In mixed phytoplankton populations the first step is to identify the target cells. This can be achieved on the basis of cell or nucleus size, e.g. the large nucleus of Prorocentrum micans makes this dinoflagellate distinguishable from accompanying species by propidium iodide staining only (Peperzak and Sandee, unpublished). Small species with body scales such as Emiliania huxleyi, can be identified by their scatter characteristics (Van Bleijswijk and Veldhuis, 1995). A combination of cell specific immunofluorescence and propidium iodide DNA staining is shown in Figure 3. Here, G. mikimotoi cells were labelled with a primary monoclonal, cell wall targeted antibody and a secondary anti-mouse antibody, conjugated with AlloPhycocyanin. The red APC-fluorescence, together with the orange propidium iodide fluorescence of the cell nuclei enabled rapid selection of the G. mikimotoi cluster.

The third example that is presented in this section, is based on observations that were made with species-specific cell surface antibodies, and with species-specific ribosomal-RNA probes (e.g. Jonker et al., 2000). Because antibodies are directed against cell surface proteins, an effect of environmental change such as nitrogen limitation, might lead to a reduction in epitopes and, in consequence, a reduction in fluorescence intensity. Fortunately, it has been shown that nutrient limitation does not change the label intensities of G. aureolum, G. nagasakiense, A. fundyense and A. tamarense (Vrieling et al., 1997; Scholin and Anderson, 1998; Anderson et al., 1999). However, the rRNA label intensity decreases considerably during nitrogen limitation (Scholin and Anderson, 1998; Anderson et al., 1999). In Figure 4 the fluorescence intensities of cells from nitrogen replete and nitrogen stressed G. mikimotoi cultures labelled with either a monoclonal antibody, or a universal rRNA probe are shown. The rRNA fluorescence decreased significantly under nitrogen stress while the antibody fluorescence remained practically unchanged. Combining the antibody with the rRNA labelling methods can therefore be used to measure the physiological state of individual phytoplankton cells in situ.
Immunofluorescent species identification + immunofluorescent fluorophore

Once a suitable immunological probe has been found to identify phytoplankton cells flow cytometrically, there are, in principle, no objections to proceed with the immunochemical labelling of certain cell components. As with rRNA labelling, the labelling of cell components can provide physiological information on a per cell basis. A variety of monoclonal and polyclonal antibodies has been produced and applied in whole cell analysis (Table 4). In most studies microscopical techniques have been used, but there are a few culture studies that have employed a flow cytometer. Buma et al. (1995) measured DNA damage from UV radiation in the diatom *Cyclotella sp.* The CO₂-fixing enzyme Rubisco was measured flow cytometrically by Orellana and Perry (1992, 1995) in the diatom *Thalassiosira weisflogii* and in the prymnesiophyte *Isochrysis galbana*. To date there are no studies in which the species and a cell component are measured simultaneously with immunological probes (cf. Jochem, 2000).

Before such combined measurements can be made it is necessary that cell surface antigenicity, or species labelling, is not reduced too much by the protocols that are used to fix and permeabilise the cells. The antigenicity of the physiological indicator, e.g. Rubisco, can also be affected by the fixation and permeabilisation treatments (Orellana and Perry, 1995). Because the cell wall has been permeabilised to allow entry of the immunological probe, care must be taken that after labelling the probe-indicator complex does not leak out of the cells. This implies that much work has to be performed to develop reliable immunological combination assays.

As becomes clear from the data in Table 1, much effort in immunological species identification has been directed towards toxic phytoplankton. Phytoplankton cells from this group, however, are not always toxic and the toxin concentration is variable. Ecophysiologial knowledge is to be gained, if the effect of the variables that determine toxicity can be measured per cell in culture experiments, as well as in field studies.

Normally, toxicity is measured in samples composed of many cells with a known cell concentra-

| Measurement | PAβ/ MAβ directed against | species | class | fcm applied | culture / field | reference |
|-------------|---------------------------|---------|-------|-------------|----------------|-----------|
| cell cycle  | MAβFITC                   | PCNAβ   | Dunaliella tertiolecta | chlorophyceae | no            | c         | Lin et al., 1995, 1996a |
|             | MAβPCNA                   | Dunaliella tertiolecta | chlorophyceae | no           | c             |           |
|             | MAβPCNA                   | Thalassiosira weisflogii | bacillariophyceae | no           | c             |           |
|             | MAβPCNA                   | Skeletonema costatum | bacillariophyceae | no           | c             |           |
|             | MAβPCNA                   | Rhizosolenia sp. | bacillariophyceae | no           | f             |           |
|             | MAβPCNA                   | Ethmodiscus rex | bacillariophyceae | no           | f             |           |
|             | PAβp34cdc2β               | Dunaliella tertiolecta | chlorophyceae | no           | c             |           |
| DNA damage  | MAβFITC                   | thymine dimers | Cyclotella sp. | bacillariophyceae | no           | c         | EPICS-CS* | Buma et al., 1995 |
| Iron limitation | PAβFITC                 | flavodoxin | Dunaliella tertiolecta | chlorophyceae | no           | c         | La Roche et al., 1993 |
| Nitrate utilization | PAβFITC         | nitrate reductase | several species | bacillariophyceae | no           | f         | Balch et al., 1988 |
| Nitrogen limitation | PAβFITC      | Rubisco | Isochrysis galbana | prymnesiophyceae | no           | c         | Falkowski et al., 1989 |
| Photosynthetic rate | PAβFITC     | Rubisco | *Thalassiosira weisflogii* | bacillariophyceae | EPICS-Profile* | c | Orellana and Perry 1992 |
| RUBISCO concentration | PAβFITC   | Rubisco | *Isochrysis galbana* | prymnesiophyceae | EPICS-Profile* | c | Orellana and Perry 1995 |
| RUBISCO distribution | PAβFITCa   | Rubisco | Dunaliella tertiolecta | chlorophyceae | no           | c         | Lin and Carpenter 1997a+b |
|             | Rubisco | Isochrysis galbana | prymnesiophyceae | no           | c             |           | Lin and Carpenter 1996a |
|             | Rubisco | *Thalassiosira weisflogii* | bacillariophyceae | no           | c             |           |
|             | Rubisco | *Thalassiosira oceanica* | bacillariophyceae | no           | c             |           |
|             | Rubisco | Skeletonema costatum | bacillariophyceae | no           | c             |           |
|             | Rubisco | *Aureococcus anophagefferens* | pelagophyceae | no           | c             |           |

*PAβ source = Falkowski et al., 1989*
*for whole cells after permeabilisation*
*best fixative was 96% EtOH*
*light harvesting Chl a / c fucoxanthin complex*
*probe reacted positively with 38 species of algae using western immunoblotting*
*fluorophore used: FITC = Fluorescein IsoThioCyanate, PE = PhycoErythrin, no = no fluorescent probe*
*microscopical tests on field samples were performed*
*Proliferating Cell Nuclear Antigen (PCNA) and p34cdc2 kinase are cell cycle proteins*
tion, and the toxin concentration per cell is averaged. The benefit of a per cell toxin assay was already recognised by Yentsch (1981) who used chemical methods to measure saxitoxin in cultured *A. tamarense* cells with a flow cytometer. Later, immunological methods were employed against saxitoxins and okadaic acid in thin sections of *Alexandrium* and *Prorocentrum* cells (Table 5).

Because antibodies can pass permeabilised cell walls, as in whole cell physiological measurements (Table 4), there is no principal objection to employ the same technique in toxin measurements. Both polyclonal and monoclonal antibodies have been prepared against the major phytotoxins (Table 5), although some of them have not been tested on toxic cells directly but on cell extracts or tissue fluids.

One of the problems in toxin analysis could be the low toxin concentration per cell. Therefore, the maximum toxin concentrations reported, for a number of species from a variety of phycotoxic classes, has been listed in Table 6. In most species the maximum toxin concentrations are in the fmol (10^-15 mol) cell^-1 range, and although this is quite low, in comparison the cellular Rubisco concentration is in the amol (10^-18 mol) range (Falkowski et al., 1989). Rubisco can be measured flow cytometrically (Orelolana and Perry, 1992, 1995), and the conclusion

### Table 5 – Immunological toxin assays. All plankton species listed belong to the dinoflagellates. The only flow cytometric assay was that of Yentsch (1981), however, no immunochemical was used (note b).

| toxin species | applied in | PAb/MAb | reference |
|---------------|------------|---------|-----------|
| *Gymnodinium breve* | cell extract | PAb | Baden et al., 1988 |
| *Gambierdiscus toxicus* | fish tissue fluid | MAb | Hokama, 1993 |
| *Dinophysis spp.*, *Prorocentrum spp.* | shellfish cell thin sections | MAb | Matsuura et al., 1994a+b |
| *G. tamarensis var. excavata* | individual cells | neither | Yentsch, 1981 |
| *G. tamarenses* | cells | PAb/FC | Anderson and Cheng, 1988 |
| *Alexandrium fundyense* | | M+PAb | Doucette and Anderson, 1993 |

### Table 6 – Maximum toxin concentrations in cells of various dinoflagellate, raphidophyte and diatom species. For comparison, the Rubisco concentration in the small prymnesiophyte *Isochrysis galbana*, measured immuno flow cytometrically is listed

| species | toxin | concentration fmol cell^-1 | reference |
|---------|-------|---------------------------|-----------|
| *Alexandrium ostenfeldii* | total PSP toxins | 2 | Hansen et al., 1992 |
| *Alexandrium tamarenses* | total PSP toxins | 40 | Boyer et al., 1987 |
| *Chattonella antiqua* | CaTx I-IV | 6 | Khan et al., 1996a |
| *Dinophysis acuminata* | DTX-I | <1 | Johansson et al., 1996 |
| *Dinophysis acuminata* | Okadaic Acid | 29 | Johansson et al., 1996 |
| *Dinophysis acuminata* | Okadaic Acid | 32 | Cembella, 1989 |
| *Dinophysis acuta* | DTX-I | 9 | Johansson et al., 1996 |
| *Dinophysis acuta* | Okadaic Acid | 17 | Johansson et al., 1996 |
| *Dinophysis norvegica* | Okadaic Acid | 41 | Cembella, 1989 |
| *Fibrocapsa japonica* | FjTx I-IV | 5 | Khan et al., 1996b |
| *Gymnodinium breve* | total brevetoxins | 12 | Roszell et al., 1990 |
| *Gymnodinium catenatum* | total PSP toxins | 200 | Oshima et al., 1993 |
| *Gymnodinium sp.* | Gymnodimine | 0.6 | Seki et al., 1996; M. Satake, pers.comm. |
| *Pseudonitzschia multiseries* | Domoic Acid | 38 | Pan et al., 1996 |
| *Pseudonitzschia seriata* | Domoic Acid | 108 | Lundholm et al., 1994 |
| *Isochrysis galbana* | RUBISCO | 80 - 390 10^-3 | Falkowski et al., 1989 |

*Gymnodinium cf. mikimotoi* (M. Satake, pers.comm.)

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**IMMUNO FLOW CYTOMETRY 177**

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seems permitted that toxins can be measured too, even in moderately toxic cells. Again, trials to establish the feasibility of such measurements have not been undertaken.

Non-immunological species identification + immunofluorescent fluorophore

This combination is in fact the same as (ii), but the species identification is now of a non-immunological nature. Examples of such species identification techniques, using a flow cytometer, include recognition of size and structure, and fluorometric labelling with dyes, lectins or RNA probes. Because most of these techniques have been reviewed in previous sections they will only be shortly discussed here.

Large dinoflagellates and small scaly prymnesiophytes can be recognised by scatter and time of flight (Van Bleijswijck and Veldhuis, 1995; Peperzak and Sandee, unpublished). Calcofluor white can be used to stain thecate dinoflagellates when the flow cytometer is equipped with a UV laser (Table 3). In certain cases fluorescent lectins may be employed (Lopez Rodas and Costas, 1997); lectins bind to specific carbohydrates on the cell wall. Lectins are not species-specific, but when a set of lectins is used, species can be specifically identified. A relatively new method in species identification employs fluorescent RNA probes (Miller and Scholin, 1998; Scholin and Anderson, 1998).

If species identification can be achieved flow cytometrically with one of these non-immunological probes, the possibilities of qualitative labelling with a second, (non)-immunological fluorophore as described in (i) and (ii) can be examined.

CONCLUDING REMARKS

The number of immunological techniques that, in principal, can be applied in flow cytometric phytoplankton research has increased in the last decade. Of these techniques the identification and counting of phytoplankton species, especially toxic ones, has received most attention. The adjustment of epifluorescence microscopy protocols for use in flow cytometry, can rapidly increase the number of species to be counted automatically. New fluorescent probes should increase label intensity and hence detectability of the target cells in mixed populations. Once the flow cytometric identification of a single cell is well established, the measurement of several ecologically relevant physiological variables relating to primary production, nutrient limitation or toxin concentration will be a next feasible goal in this field of research.

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