Oxidative effects of glyphosate on the lipophobic intracellular environment in the microalgae Chlorella vulgaris

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Abstract: The studied hypothesis is that the herbicide glyphosate (GLY) can affect the oxidative balance in the lipophobic intracellular medium in non-target Chlorella vulgaris cells. Analytical GLY and RoundUp (RUP) supplementation, affected the growth profile. A significant 42% decrease in the cellular biomass in stationary (St) phase was observed in cultures supplemented with either 5 µM of GLY or RUP, as compared to control cultures. The treatment with 0.3 µM of GLY generated non-significant effects on the oxidation rate of 2', 7' dichlorofluorescein diacetate (DCFH-DA), neither in exponential (Exp) nor in St phase of development, as compared to control cultures. However, the treatment with either 5 µM GLY or 0.3 and 5 µM RUP lead to a significant decrease in the DCFH-DA oxidation rate, as compared to control cultures. The lipid radical (LR•) generation rate, detected by Paramagnetic Resonance Spectroscopy (EPR), was significantly increased in the presence of RUP, in Lag and Exp phase of growth. The non-enzymatic antioxidants, α-Tocopherol (α-T) and β-Carotene (β-C), are aimed to protect membranes against the damage produced by the radical reactions. The content of β-C was not significantly affected, as compared to control cultures, by any of the treatments, in both growth phases of cellular development. The content of α-T was significantly decreased by the supplementation with either 0.3 or 5 µM of RUP or 5 µM GLY. The LR•/α-T ratio, used as indicator of the oxidative balance in the lipophobic cellular media, was significantly different between samples obtained from control and RUP-exposed microalgae in both, Exp and St phase of development, with either 0.3 or 5 µM RUP. The data presented here showed evidence that suggested that oxidative balance in the lipophobic environment was affected by either GLY or RUP.

Introduction

Glyphosate (C6H16NO5PS) (GLY) is a broad spectra herbicide that is mainly used for weed control in crops of agricultural interest. Studies over the past decade in a variety of biological models, have shown that GLY and its commercial formulations are responsible for associated toxicity of variable magnitude. GLY enters aquatic ecosystems by accidental spraying, drifts or surface runoff (Altamar Ríos, 2007). GLY generates the death of the target systems by affecting their ability to synthesize proteins essential for survival. The activity of the enzyme 5-enolpiruvilshikimato-3-phosphate synthase, which is part of the metabolic pathway for production of aromatic amino acids, is inhibited by the GLY that behaves as an analogue of the second substrate (phosphoenolpyruvate). Moreover, there is growing evidence that indicate that GLY (and its formulations) produce oxidative stress in cyanobacteria (Chen et al., 2012), microalgae (Lipok et al., 2010) and higher plants not natural targets for this herbicide (Ahsan et al., 2008). Also, it has been suggested that the pathological effects in humans are due to a mechanism of free radical production due to the nonspecific chronic inflammation that it generates (Altamar Ríos, 2007). In the air, GLY is eliminated through reactions with hydroxyl radical (•OH), and in the soil, it is degraded by the action of microorganisms, producing methylphosphonic ammonium acid (AMPA) and carbon dioxide (CO2) (Altamar Ríos, 2007). In these reactions, free radicals and recombinant or fibrogenic cytokines are eliminated (Altamar Ríos, 2007). However, the real dimension of the magnitude of the oxidative effect generated by exposure to the herbicide is still a matter of deliberation.

The exposure of cultures of the algae Chlorella kessleri, to GLY at concentrations of 50 mg l⁻¹, much higher than the
concentrations usually found in the environment, triggers oxidative stress, evidenced by the inhibition of its growth and morphological damage with disorganization of chloroplasts, lipid damage, increased protein concentration, and increased antioxidant defenses (GSH; superoxide dismutase, SOD and catalase, CAT) (Romero et al., 2011). Choi et al. (2012) found that exposure to GLY affected chlorophyll-a fluorescence in five microalgal species, with certain interspecific differences in GLY sensitivity. Gomes and Juneau (2016), and Gomes et al. (2016a, 2016b) showed a strict relationship between GLY-induced oxidative stress and its deleterious effects on the photosynthetic processes. Also, Eker et al. (2006) and Cakmak et al. (2009) have reported that GLY can bind to different essential metals, causing a decrease in its bioavailability. Duke et al. (2012) have pointed out some contradictions on the mechanism of GLY to bind. The microalga *Chlorella vulgaris*, is frequently used for analysis of toxicity since it is easily grown under laboratory conditions. This alga is found in many freshwater aquatic environments, and it is seen as a good bioindicator of the presence of environmental contaminants. Qian et al. (2009) and Reno et al. (2016) have reported several studies on effects of the presence of contaminants, such as GLY and its commercial products, on *C. vulgaris*. More recently, Ostera et al. (2020) reported experimental evidence that strongly suggested that oxidative balance in the hydrophilic cellular environment is affected by GLY independently of the preparation used to administrate the product, even at low-medium concentrations currently in used. Moreover, it seems that the effect is reversible, either due to the magnitude of the herbicide-dependent damage or by the antioxidant activity endogenously activated. However, the reactive oxygen radical (ROS) and nitrogen radical species lead to damage to macromolecules, not only in the hydrophilic but in the hydrophobic medium, as well. The lipid peroxidation is understood as the oxidative deterioration of the polyunsaturated lipids and conduct to effects such as alterations in the structure, fluidity and permeability of cellular membranes. The term lipid radicals (LR•), generated by peroxidation, include alkyl, peroxyl (ROO•) and alkoxy radicals. Non-enzymatic lipid soluble antioxidants, such as α-Tocopherol (α-T) and β-Carotene (β-C), are aimed to protect, repair, and remove of the radical species generated. The LR•/α-T ratio was established as an appropriate indicator of oxidative condition in the hydrophobic cellular space (Malanga et al., 2009). Considering the increasing awareness of the environmental risks generated by the use of GLY and its commercial formulations, the cellular pathways that could be affected in non-target organism by the exposure to the herbicide should be further studied. The hypothesis of the present work is that oxidative alterations could affect *C. vulgaris* redox balance in the hydrophobic cellular medium, consuming lipid soluble antioxidants, after the exposure to either pure GLY or the commercial formulate RUP. The effect of GLY and its formulate was measured on the biomass and rate of growth of the cultures, the oxidation rate of DCFH-DA, the generation rate of LR•, and the content of lipid soluble antioxidants (α-T and β-C).

### Materials and Methods

#### Culture conditions and experimental design

Stock axenic cultures of *Chlorella vulgaris* CPCC90 (Origin: Canadian Phycological Culture Centre, strain identification: C90) were grown in sterilized Bold’s Basal medium supplemented with 1 g l⁻¹ glucose (Bold and Wynne, 1978). Cells were grown at 20 ± 1°C under light/dark cycles of 12:12 h. The irradiance at the surface of the culture was approximately 38 W m⁻¹ of photosynthetically active radiation (PAR, Philips 50-W day-light-fluorescent light). Cell abundance was monitored by spectrophotometry (A = 600 nm) and cell counting using a hemocytometer. For cell counting, samples were taken and kept in plastic tubes with a solution of formaldehyde in water 10% (v/v).

Cells were grown in Bold’s Basal medium with 1 g l⁻¹ of glucose (control cells). For the treatments, the media was supplemented with either analytical grade GLY (Sigma-Aldrich) or a GLY-based herbicide (RoundUp UltraMax®, RUP, donated by colleagues) to a final concentration of active ingredient (GLY) of 0.3 µM or 5 µM (50 µg l⁻¹ and 800 µg l⁻¹, respectively). All determinations were done in exponential (Exp) (day 7) and/or stationary (St) (day 14) growth phases. In each phase the cells were harvested by centrifugation at 9,300 g at 4°C. Cells were preserved at −80°C until the time of processing.

The diagram in Fig. 1 shows the experimental protocol designed for these studies.

#### Oxidation rate of DCFH-DA

Cellular cultures containing at least 2 × 106 cells ml⁻¹ were used for assaying the oxidation rate of the dye. The harvested cells were resuspended using 400 µl of a Tris–HCl 40 mM pH 7.4 buffer. Two subsamples of this suspension containing 200 µl each were taken and transferred to new plastic tubes, then centrifuged at 9,280 g at 4°C for 10 min. The supernatant was discarded and 100 µl of the same buffer were added to the pellet and resuspended. This new suspension was sonicated for 30 s using 10 s continuous pulses. An aliquot of 96 µl of buffer and 4 µl of DCFH-DA prepared using the same proportion of buffer:DCFH-DA as the samples and their blanks. Both, the samples and the blanks, were incubated for 45 min at 30°C, followed by centrifugation at 9,280 g for 5 min at 4°C. Supernatants were transferred to a 96-wells plaque. Fluorescence was determined at λex = 488 nm and λem = 525 nm using a Varioscan Lux microplate reader.

#### Generation rate of LR•

The generation rate of LR• was determined by electronic paramagnetic resonance (EPR) using α-(4-pyridyl N-oxide)-N-tert-butylnitrone (POBN) as a spin trap. Harvested cells were resuspended in 400 µl of a 0.5 M KH₂PO₄/K₂HPO₄ pH 7.0 buffer and sonicated for 15 s. An aliquot of 200 µl of the resulting homogenate was mixed with an equal volume of a 0.5 M solution of POBN prepared in the same buffer. Samples were placed in a glass micropipette and incubated...
at 30°C for 30 min before LR• determination. Spectra were measured using a Bruker (Karlsruhe, Germany) spectrometer ECS 106 with a cavity ER 4102ST, operating with the following instrument settings: 18°C room temperature, 50 kHz modulation frequency, 20 mW microwave power, 1 G modulation amplitude, 81.92 ms time constant and $2 \times 10^4$ receiver gain (Buettner and Jurkiewicz, 1993). Quantification of the spin adduct was performed using an aqueous solution of 4-hydroxy-2,2,6,6-tetramethylpiperidinyloxy (TEMPOL) introduced into the same sample cell and subjected to the same treatment as the samples. EPR spectra for both sample and TEMPOL solutions were recorded at exactly the same spectrometer settings and the first derivative EPR spectra were double integrated to obtain the area intensity, from which the content of the radical was calculated according to Kotake et al. (1996).

Content of lipid soluble antioxidants
The content of α-T and β-C was determined using high performance liquid chromatography (HPLC) with electrochemical detection in isocratic conditions according to Desai (1984) with modifications. Harvested cells were resuspended in 250 µl of a 1:1 methanol:ethanol solution and sonicated for 1 min for the extraction of the antioxidants. Afterwards, samples were kept for 30 min in the dark at 4°C. This procedure was repeated after the addition of 1 ml of hexane. After the last 30 min of incubation with hexane, samples were centrifuged at 2,320 g for 10 min at 4°C. An aliquot of 750 µl of the organic phase was then transferred to a new plastic tube and evaporated to dryness using gaseous N2 at room temperature. This extract was kept at −80°C until the time of injection. The HPLC procedure was performed using a Perkin-Elmer LC Binary Pump 250, ESA Coulochem II detector, ESA Model 5011 analytical cell, ESA Model 5020 Guard Cell, Perkin Elmer Model 1022 integrator and a manual injector with a 20 µl loop. The column employed was a Supelcosil LC-8 (150 mm × 4.6 mm; pore size 3 µm). Methanol and a 20 mM solution of lithium perchlorate in distilled water in a 97:3 proportion was used as the mobile phase. Flux was set to 1 ml min-1 for a total run time of 13 to 15 min. Samples were resuspended using 150 µl of a 1:1 methanol:ethanol mixture and filtered using Nanosep MF 0.2 µm filters previous to injection. Quantification of antioxidants was made by injecting solutions of analytical standards of known concentration.

Statistical analysis
All statistical analyses were performed using the software GraphPad Prism 6.0. The results are expressed as mean ± ESM of at least three independent experiments, with three replicates in each experiment and were analyzed using 1-way or 2-way ANOVA with $p = 0.05$ or $p = 0.1$, unless otherwise specified.

Results
The growth cycle of C. vulgaris microalgae comprises three phases: lag phase (days 0–4), Exp phase (days 4–0) and St phase (days 11–14). To evaluate the effects of GLY and

**FIGURE 1.** Diagram illustrating the experimental design used in this work.
RUP treatment on the biomass of the cultures, the cell count of the cultures was followed over 14 days, as shown in Fig. 2. A significant 42% decrease in the cellular biomass in St phase was observed in cultures supplemented with either 5 µM of GLY or RUP, as compared to control cultures (Table 1).

Growth rate (µ) was calculated according to Tajnaiová et al. (2020). The growth rate of cultures exposed to either 5 µM of GLY or RUP in Exp phase of development showed a non-significant decrease, as compared to control cultures (0.16 ± 0.04, 0.4 ± 0.1 and 0.7 ± 0.2 10⁷ cells ml⁻¹ day⁻¹, respectively). Biomass was not affected after exposure of the cells to either 5 µM of GLY or RUP in Exp phase of development in agreement with previous reports (Ostera et al. 2020); however, a significant decrease in biomass, evaluated at day 14 in cultures treated with either GLY or RUP, was observed as compared to control cultures (Table 1). Fig. 3 shows the DCFH-DA oxidation rate in control and GLY- and RUP-treated cultures in Exp and St growth phase. The treatment with 0.3 µM of GLY generated non-significant effects on the oxidation rate of DCFH-DA, neither in Exp nor in St phase of development as compared to control cultures. However, the treatment with GLY 5 µM lead to a significant decrease (62%) in the DCFH-DA oxidation rate as compared to control cultures either in Exp and St phase of cellular development, respectively. C. vulgaris cultures showed a significant decrease in the DCFH-DA oxidation rate after supplementation with either 0.3 µM or 5 µM of RUP in both phases of development. In cultures treated with 0.3 µM of RUP a significant 45% and 56% decrease in the oxidation rate was observed in Exp and St phase as compared to control cells, respectively. In the presence of 5 µM RUP, a significant decrease in the oxidation rate of the dye was observed in Exp and St phase compared to control cultures (52, 47%, respectively).

The LR• generation rate for control and GLY- or RUP-treated cultures is shown in Fig. 4. The LR• generation rate in Exp growth phase was not significantly affected by the supplementation of cultures with either 0.3 µM of GLY or RUP. However, a significant increase (23%) in the oxidation rate was observed only in cultures supplemented with 5 µM RUP, as compared to control in St phase.

Data in Table 2 shows the content of lipophilic antioxidants, α-T and β-C, in control, GLY- and RUP-treated cultures in both phases of development. The content of β-C was not significantly affected, as compared to control cultures, by any of the treatments in both growth phases of cellular development. In Exp phase, the content of α-T was significantly affected by the supplementation with 0.3 µM and 5 µM of RUP, showing an 87 and 84% decrease, as compared to control cultures, respectively. In St growth phase supplementation with either 0.3 or 5 µM GLY generated a significant 28 and 65% decrease in the α-T content, as compared to control cultures in the same growth phase, respectively. A significant 92% decrease was determined in the α-T content in cultures treated with either 0.3 or 5 µM of RUP, as compared to control cultures in this growth phase.

The LR•/α-T content ratio, considered as an indicator of the balance between free radical damage and antioxidant protection in the hydrophobic medium (Malanga et al. 2009), was significantly different between samples obtained from control and RUP-exposed microalgae in both, Exp and St phase of development at either 0.3 or 5 µM formulate (Fig. 5). However, the ratio was not affected when C. vulgaris cells were exposed to 0.3 µM GLY in the Exp phase of development (Fig. 5 Insert), showing a more drastic effect of RUP, as compared to GLY tested under the same experimental conditions.

**Discussion**

GLY can affect non-target organisms in the aquatic environment, including microalgae and cyanobacteria, which share photosynthetic routes similar to those described for terrestrial plants (Hernández-García and Martínez-Jerónimo, 2020). They are not only affected in a similar fashion the shikimate pathway, but they also experience oxidative stress as well (Ostera et al., 2016). It was suggested than in addition to producing physiological and metabolic damage that impairs algal growth (Wu et al., 2016), GLY disrupts the integrity of the cell membrane (Amorós et al., 2007). A recent report (Hernández-García and Martínez-Jerónimo, 2020) evaluated the growth of an experimental community of chlorophycean microalgae (Ankistrodesmus falcatus, C. vulgaris, Pseudokirchneriella subcapitata, and Scenedesmus incrassatus), by comparing their growth rate in the presence of the herbicide Faena" (1.564 mg GLY/l

![FIGURE 2. Kinetic of the biomass of the cultures of C. vulgaris. Biomass of Control cells (○); and cells grown in the presence of GLY 0.3 µM (□); GLY 5 µM (△); RUP 0.3 µM (▲); RUP 5 µM (▲).](image)

**TABLE 1**

| Treatment      | Biomass (10⁷ cells ml⁻¹) |
|----------------|-------------------------|
|                | Exp phase               | St phase               |
| Control        | 1.3 ± 0.2               | 3.1 ± 0.2              |
| GLY 0.3 µM     | 1.4 ± 0.3               | 3.7 ± 0.2              |
| GLY 5 µM       | 1.1 ± 0.2               | 1.8 ± 0.2*             |
| RUP 0.3 µM     | 1.5 ± 0.2               | 3.4 ± 0.2              |
| RUP 5 µM       | 1.2 ± 0.2               | 1.8 ± 0.1*             |

Note: *significantly different from control cultures (one-way ANOVA, p < 0.05).
during 14 days for *C. vulgaris*). In that study, negative, significant effects on the growth and physiology of green microalgae (revealed through the assessment of macromolecular and enzymatic biomarkers) were documented. In the same direction, the results reported here employing GLY (either pure or as RUP formulate), in a lower dose than the IC40 used previously, is in agreement with those reports since biomass of the cultures of *C. vulgaris* was significantly decreased in St phase of development. The observed effect in microalgae may result from the inhibition of the enzyme 5-enolpyruvylshikimate-3-phosphate synthase (EPSPS) in the shikimate pathway (Giesy et al., 2000; Pérez et al., 2011) and the possible oxidative stress triggered by the exposure to the herbicide, causing an imbalance in the enzymatic response (Ostera et al., 2016).

The oxidation rate of DCFH-DA was successfully used as a general index of ROS generation in algae exposed to UV-B radiation (Malanga and Puntarulo, 1995), and in soybean roots in vivo exposed to Fe overload, and as an in vitro assay to study the ability of isolated microsomal membranes to generate superoxide anion (O$_2^-$) in the presence of NADPH (Caro and Puntarulo, 1996). Even more, Patetsini et al. (2013) examined, employing flow cytometry, the influence of environmentally relevant concentrations of two pesticides (chlorpyrifos and penoxsulam) on mussel physiological status. These works established this methodology as a reliable biomarker for the evaluation of pollution or other environmental stressors. The data reported here indicated that after the exposure to a low GLY concentration (0.3 μM) the oxidation rate of DCFH-DA was not affected neither in Exp nor in St phase of development of *C. vulgaris* cells. At

| Exp phase | α-T content (pmol 10$^{-4}$ cells) | β-C content (pmol 10$^{-4}$ cells) |
|-----------|---------------------------------|---------------------------------|
| Control   | 0.4 ± 0.1                        | 0.05 ± 0.02                     |
| GLY 0.3 μM| 0.3 ± 0.2                        | 0.04 ± 0.03                     |
| GLY 5 μM  | 0.19 ± 0.03                      | 0.03 ± 0.02                     |
| RUP 0.3 μM| 0.05 ± 0.02**                    | 0.006 ± 0.003                   |
| RUP 5 μM  | 0.061 ± 0.004*                   | 0.012 ± 0.004                   |

| St phase  | α-T content (pmol 10$^{-4}$ cells) | β-C content (pmol 10$^{-4}$ cells) |
|-----------|---------------------------------|---------------------------------|
| Control   | 0.75 ± 0.07                     | 0.04 ± 0.02                     |
| GLY 0.3 μM| 0.54 ± 0.05**                   | 0.03 ± 0.01                     |
| GLY 5 μM  | 0.26 ± 0.08*                    | 0.03 ± 0.01                     |
| RUP 0.3 μM| 0.06 ± 0.01**                   | 0.005 ± 0.002                   |
| RUP 5 μM  | 0.06 ± 0.01*                    | 0.005 ± 0.002                   |

Note: *significantly different from control cultures in the same growth phase (one-way ANOVA, *p* ≤ 0.05). **significantly different from control cultures in the same growth phase (one-way ANOVA, *p* ≤ 0.1).
the sarcosine path, it can react with the phosphonate group of GLY (Borggaard and Gimsing, 2008). Fe oxides in soils due to its high affinity with the number of dissolved Fe ions. Fe oxides are the GLY scavengers in the magnetite/UV system was ascribed to a high inorganic ions. Overall, an effective photo-degradation of ANOVA, p

![Figure 5](image)

**FIGURE 5.** LR\(\alpha\)-T ratio in Control and GLY/RUP treated cultures in Exp (inset) and St (main figure) growth phase. *significantly different from control cultures in the same growth phase (one-way ANOVA, p < 0.05).**

The sequestration of GLY on Fe oxides has been extensively studied, but the knowledge on the subsequent GLY degradation on Fe oxides is limited. Due to the extensive existence of goethite and magnetite in nature, it should have a great impact on the fate of pesticides and other organic pollutants in the environment. As ROS can be formed in situ on a mineral surface, it can be reasonably inferred that a similar degradation reaction, described in the environment, will occur in the cells. Thus, the data reported here showing that the amount of reactive ROS available to react with the dye was decreased in the presence of GLY, as compared with the magnitude in its absence, could suggest that similar reactions to those reported in the cellular environment could occur in the cells, and that the presence of Fe is key factor in this scenario. However, until now this mechanism was not studied in biological organisms and further research in this field is strongly required.

(b) Lipid soluble antioxidants are the best candidates to be involved in the protection mechanisms against GLY-dependent damage since oxidative balance in the hydrophilic medium of C vulgaris cells was affected by this treatment (Ostera et al., 2020). The data reported here clearly suggested that oxidative redox balance was altered at the lipophilic cellular media, such as indicated by the increase in the LR\(\alpha\) generation rate and in the LR\(\alpha\)-T ratio, after exposure of the microalgae to RUP.

The content of \(\beta\)-C, that is very low in the cell as compared to other antioxidants, did not change by GLY administration. Tocopherol biosynthesis involves two metabolic pathways: the first is via homogentisic acid through the cytosolic shikimate pathway whereas the second is the plastid methylerthritol phosphate pathway for the tocopherol tail (phytyle diphosphate) synthesis. Finally, the assembly of the aromatic ring and phytyle diphosphate tail gives rise to tocopherols (Herrmann and Weaver, 1999). The significant decrease in the content of \(\alpha\)-T observed in C. vulgaris cells supplemented with either GLY or RUP, strongly suggested that the consumption of this antioxidant is an important defense barrier to the deleterious effect of the herbicide. On the other hand, \(\alpha\)-T works synergistically with other antioxidant species such as carotenoids, glutathione (GSH), and ascorbate (AH-) and preserve plastid redox homeostasis in stressed plants (Munné-Bosch, 2005). The formation of these conjugated low molecular weight antioxidants protects further plant metabolism against stress injuries (Szarka et al., 2012). There are two possible means of enhancing conjugated antioxidant formation, increasing either the biosynthesis or the redox cycling of these compounds, and both can result in an increased biological activity of the compounds. In the cell, the interdependence of these antioxidants could play a basic role in electron flow (Gill and Tuteja, 2010).

Synthesis of various aromatic amino acids and precursors of numerous secondary metabolites, like vitamins and pigments takes place through the shikimate pathway (Herrmann and Weaver, 1999). In this pathway, one of the glycolytic cycle intermediates, phosphoenolpyruvic acid, and another takes place through the shikimate pathway. The reversible synthesis of 5-enolpyruvylshikimate 3-phosphate and phosphoenolpyruvate is mediated by the enzyme 5-enolpyruvylshikimate 3-phosphate synthase (Velske and Cejpek, 2006) and inhibited by GLY (Lushchak and Semchuk, 2012). Thus, it can be postulated that the highest GLY concentration (5 \(\mu\)M) the oxidation rate of the dye was significantly decreased. This decrease was also detected after treatment with RUP at the studied concentrations. This unexpected response could be due to different possible effects of the herbicide:

(a) GLY that could be degraded by processes, such as metal oxides/UV, Fenton/UV, Fenton-like/UV, ozonation, chlorination (Yang et al., 2018) was described to split into PO43- and sarcosine by the chemisorption of H2O and OH in the existence of Fe oxides (Jaisi et al., 2016). For the sarcosine path, it can react with OH and O2- to form glycine and formaldehyde directly. Formaldehyde would be finally oxidized to formic acid and mineralized to CO2 and water. Meanwhile, the decarboxylation of glycine might lead to the formation of methylamine due to the existence of OH (Hidaka et al., 1997). Through the oxidation of OH and O2- methylamine could be further oxidized to ammonium (NH4+), nitrite (NO2-) and nitrate (NO3-) (Manassero et al., 2010). EPR studies verified the existence of OH and O2- in both goethite/UV and magnetite/UV systems, and OH dominated the photo-degradation of GLY (Yang et al., 2018). In addition, as a highly reactive oxidizing agent, OH and O2- could oxidize GLY to amino acid and carboxylic acid, and finally mineralized to inorganic ions. Overall, an effective photo-degradation of GLY in the magnetite/UV system was ascribed to a high capacity of generating ROS, crucially caused by a larger number of dissolved Fe ions. Fe oxides are the GLY scavengers in soils due to its high affinity with the phosphonate group of GLY (Borggaard and Gimsing, 2008). The sequestration of GLY on Fe oxides has been extensively studied, but the knowledge on the subsequent GLY degradation on Fe oxides is limited. Due to the extensive existence of goethite and magnetite in nature, it should have a great impact on the fate of pesticides and other organic pollutants in the environment. As ROS can be formed in situ on a mineral surface, it can be reasonably inferred that a similar degradation reaction, described in the environment, will occur in the cells.
the more drastic decrease in α-T content observed in the St phase. as compared to Exp phase, could be due to this inhibitory effect of GLY in the synthesis of the antioxidant required as the growth of the culture proceed.

(c) A contribution of the already mentioned effects (a) and (b) to the decrease of the oxidation rate of the dye in cells exposed to either GLY or RUP.

The abundant and controversial studies on GLY effect on photosynthetic organisms is a clear indication of the complexity of the scenario faced when oxidative metabolism, in both target and non-target plants, is characterized. Further analysis is still required to solve this point. The data presented here are experimental evidence of the contribution of GLY-dependent effects on redox balance in the lipophilic cellular media, that is reflected in the consumption of endogenous antioxidant α-T leading to a potential decrease of the cells ability to afford further oxidative challenges. Moreover, this study shows that the formulate RUP is able to damage the cellular membranes causing deleterious effects on the cells. This is a critical issue for the agrochemical use of the herbicide, since the addition of surfactants is very much used to enhance the efficiency of the product. Thus, characterization of the RUP effects will be encouraged to provide a tool to find new alternatives for its use and, to assess the real contribution of GLY to the reported effects after exposure in the natural environment.

Availability of Data and Materials: All data generated or analyzed during this study are included in this published article (and its supplementary information files).

Author Contribution: GM and SP participated in the designing of the original idea and in the manuscript writing. JMO contributed to experimental work, the bibliographic search and in the manuscript writing. All authors have approved the final article.

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