A comparative study of heavy metal accumulation and antioxidant responses in *Vaccinium myrtillus* L. leaves in polluted and non-polluted areas

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**Abstract** The purpose of this study was to explore a possible relationship between the availability of metals in soil (Cd, Fe, Mn, Pb and Zn) and their concentrations in leaves of *Vaccinium myrtillus* L. as a species which has been reported to be a successful colonist of acid-and-heavy metal-contaminated soil. Analysis also concerned the antioxidant response of plants from three heavily polluted (immediate vicinity of: zinc smelter, iron smelter and power plant) and three relatively clean sites (nature reserve, ecological site and unprotected natural forest community) in southern Poland. The contents of glutathione, non-protein thiols, protein, proline and activity of guaiacol peroxidase in leaves of bilberry were measured. Generally, the concentrations of metals in the HNO₃ and CaCl₂ extractants of the soil from the polluted sites were higher. Moreover, the antioxidant responses were also elevated in bilberries in the polluted sites. Significant positive relationships between Cd, Pb and Zn concentrations in soil and in the plants were found. In the leaves of *V. myrtillus* from the polluted sites, higher concentrations of Cd, Pb and Zn were noted (in Miasteczko Śląskie respectively 6.26, 157.09 and 207.17 mgkg⁻¹ d.w.). We found a positive correlation between the increase in the NPTs and protein contents as well as the Cd, Pb and Zn concentrations in *V. myrtillus*. Cd, Pb and Zn also decreased guaiacol peroxidase activity. However, the activity of this enzyme increased under Fe. A decreasing trend in glutathione contents was observed with increasing iron and manganese concentrations in bilberry leaves. Parameters such as protein, non-protein –SH groups and changes in GPX activity seem to be universal, sensitive and correlated well with heavy metal stress.

**Keywords** Antioxidant response · Heavy metal · *Vaccinium myrtillus*

**Abbreviations**
NPTs Non-protein thiols
GSHt Glutathione total
GPX Guaiacol peroxidase

**Introduction**
*Vaccinium myrtillus* L. (bilberry) is a dwarf deciduous shrub which dominates plant life forms in the herbaceous layer of pine forests in temperate climates (Bialońska et al. 2007; Martz et al. 2010). It is known to play an important role in natural succession and in regulating the nutrient fluxes of forest ecosystems and thus in their productivity (Mróz and Demczuk 2010). Bilberry possesses a relatively high resistance to pollution, and it has been reported to be a successful colonist of acid-and-heavy metal-contaminated soils (Uhlig and Junttila 2001; Mróz and Demczuk 2010). In view of this fact, *V. myrtillus* L. is common both in unpolluted and polluted habitats. The accumulation of elements in its foliage has been widely used in environmental monitoring (Reimann at al. 2001; Salemaa et al. 2004; Brekken and Steinnes 2004; Bialońska et al. 2007).

Heavy metals are highly toxic to plants. Their uptake and accumulation by plant tissues cause various morphological,
physiological and biochemical responses (Doganlar and Atmaca 2011). Some metal ions are likely to remain in the cytoplasm and induce oxidative stress via generation of reactive oxygen species (ROS) such as hydrogen peroxide (H₂O₂), superoxide and hydroxyl radicals which hinder cell metabolism and lead to multiple toxic effects like lipid peroxidation, protein cleavage or DNA damages (Prasad 1999; Unyayar et al. 2006; Dazy et al. 2008; Pongrac et al. 2009). Once formed, ROS must be detoxified as efficiently as possible to minimize damage. Antioxidant systems in plants are complex and involve an array of non-enzymatic and enzymatic mechanisms capable of preventing the cascades of uncontrolled oxidation (Gratão et al. 2005, 2008).

Glutathione is one of the crucial plant metabolites in terms of intracellular defense against ROS-induced oxidative damage (Gill and Tuteja 2010). Glutathione creates complexes with heavy metals, and an induction of glutathione as well as cysteine synthesis has been documented in plants as a response to heavy metal stress. The changes in glutathione total (GSHt) level are dependent on the metal treatment and the part of plant (Arya et al. 2008; Nadgórska-Socha et al. 2012). Glutathione (GSH) is also the predominant non-protein thiol, redox buffer, phytochelatin precursor and substrate for keeping the ascorbate in reduced form in the ascorbate glutathione pathway (Demirevska-Kepova et al. 2004). Non-protein compounds rich in –SH groups (e.g. phytochelatins) are involved in metal detoxification and/or metal allocation between different organs of the plant because their main task is binding of metal ions and forming non-toxic complexes with a metal which are transported from the cytoplasm into the vacuole (Andrade et al. 2010; Yadav 2010).

Among amino acids, proline has been shown to have an important role (Sharma and Dietz 2006). Metal-induced proline accumulation has been observed, and it has been suggested that this amino acid acts as a radical scavenger, or it is involved in metal ion chelation (Andrade et al. 2009). In plants, proline constitutes <5 % of the total pool of free amino acids under normal conditions and represents 80 % of total amino acid pool under stress (Kumar et al. 2010).

Peroxidases are antioxidant enzymes which are significant in plant growth and development. Activities of these enzymes are changed under both biotic and abiotic stress conditions and are used as a potential indicator of metal toxicity (Radočić et al. 2000; Macfarlane and Burchett 2001; Baycu et al. 2006; Doganlar and Atmaca 2011). Antioxidant systems in plants may be used as early indicators of environmental stress on a target organism before morphological or ultrastructural damage occurs. Moreover, they can also be used as warning indicators of the surrounding ecosystem (Bialońska et al. 2007).

The objectives of the present study were to establish the concentrations of heavy metals as well as to determine and compare the levels of antioxidants (non-protein thiols, glutathione and proline), antioxidant enzyme (guaiacol peroxidase) and proteins content in leaves of V. myrtillus L. growing naturally at polluted and potentially unpolluted areas. We also studied the potential availability of metals (Cd, Pb, Zn, Fe and Mn) in the soil originating from the same sites. Comparing analytical results of the heavy metals and ecophysiological changes in bilberry leaves, the following questions will be investigated:

1. Do investigated physiological parameters in bilberry plants from the polluted sites vary from the ones collected at non-contaminated sites?
2. Do ecophysiological parameters are good indicators of stress caused by heavy metals in plants living in field conditions?

Material and methods

Study area

The study was performed in typical pine forests located in three heavily polluted (immediate vicinity of: the zinc smelter, iron smelter and power plant) and three relatively clean sites (nature reserve, ecological site and unprotected natural forest community). All of the sites were situated in southern parts of Poland in the Śląskie or Małopolskie provinces (Table 1 and Fig. 1).

Sample collection

Consideration of seasonal variations is essential during biomonitoring programs (Oliva et al. 2012). Metal accumulation may depend on seasonal variation. Some publications report the highest foliar levels in spring and the lowest during winter, whereas others have indicated the highest metal contents during autumn and relatively low levels during spring (Martin and Couphtrey 1982; Kim and Fergusson 1994; Brekken and Steinnes 2004; Baycu et al. 2006; Deram et al. 2006). Use of biomarkers, such as our antioxidants in biomonitoring studies, is often complicated because levels of chemical pollutants in the environment often display wide seasonal variation in response to climate and other factors. That is why we decided to sample the material three times in order to minimize the effect of seasonal factors and show changes in antioxidant response induced by heavy metals during a growing season including the most important periods of the bilberry growth (flowering stage, fructification and end of growing season).

V. myrtillus L. leaves and soil samples were collected in May, July and September of 2009. Each sampling site consisted of 25×25-m squares within which leaf and soil samples were randomly collected. The fully matured, undamaged leaves were detached from at least 20 different shrubs and
pooled into one sample per site. After collection, samples were covered with plastic bags, deposited in ice, immediately transported to the laboratory and then frozen until analysis. Soil sample were taken in the neighbourhood of the samples shrubs from a depth of 0–10 cm. At each site, soil sub-samples were combined in a composite sample.

![Location map of sampling sites](image)

### Table 1 Site description

| Sampling site no. | Sites | Latitude | Longitude | Selected soil parameters (Kandziora-Ciupa unpublished data) |
|-------------------|-------|----------|-----------|-----------------------------------------------------------|
| Polluted          |       |          |           |                                                           |
| M1                | Nearest vicinity of zinc smelter “Miasteczko Śląskie” in Miasteczko Śląskie (activities since 1968) | 50°31’22.655”N | 18°56’8.699”E | The contents of organic matter and C/N ratio are the lowest among all the sites, 7.10 % and 12.43, respectively. The pH is ∼4.45; Ca, 346.34 mgkg⁻¹; Mg, 32.33 mgkg⁻¹. |
| L2                | Nearest vicinity of iron smelter “ArcelorMittal Poland S.A.” in Dąbrowa Górnicza-Losień (activities since 1976) | 50°22’0.768”N | 19°19’22.979”E | The soil is richer in Ca and Mg, and pH is the highest compared with other sites (Ca, 3,166.17 mgkg⁻¹, Mg, 165.88 mgkg⁻¹, pH ∼4.97). |
| J3                | Nearest vicinity of power plant “Jaworzno III” in Jaworzno (activities since 1979) | 50°12’14.111”N | 19°11’51.792”E | pH ∼4.30; organic matter, 26.48 %; C/N, 19.59; Ca, 2,552.34 mgkg⁻¹; Mg, 110.59 mgkg⁻¹. |
| Unpolluted        |       |          |           |                                                           |
| P4                | Nature reserve “Pazurek” in Jaroszowiec Olkuski | 50°19’59”N | 19°37’24”E | The concentration of Mg is lowest among all sites—25.47 mgkg⁻¹; pH ∼3.67; organic matter, 15.75 %; C/N, 17.24; Ca, 587.62 mgkg⁻¹. |
| PB5               | Ecological site “Plone Bagno” in Katowice | 50°10’07”N | 19°05’18”E | The organic matter content and C/N ratio are the highest compared with other sites (36.54 % and 20.59, respectively). pH ∼3.46; Ca, 443.30 mg kg⁻¹; Mg, 65.94 mgkg⁻¹. |
| K6                | Unprotected natural forest community in Kobiór | 50°4’22.08”N | 18°56’16.656”E | The pH (3.45) of the soil and concentration of Ca (308.82 mgkg⁻¹) are the lowest among all sites. Organic matter, 22.26 %; C/N, 16.86; Mg, 50.91 mgkg⁻¹. |
Analysis of metal concentration in the soil and samples of plants

The concentrations of Cd, Pb, Zn, Fe and Mn were analysed in particular soil fractions and in the leaves of *V. myrtillus* L. The metal content in soil was estimated according to the method of Bouwmann et al. (2001) and Ostrowska et al. (1991) in the air-dried soil samples, which were sieved through a 1-mm sieve. Metals were extracted from soil with 0.01 M CaCl₂ (potentially bioavailable elements) or with 2 M HNO₃ (acid extracted elements). For the CaCl₂ extraction, 5 g of soil with 50 ml of 0.01 M CaCl₂ solution was mechanically shaken for 5 h. The HNO₃-extractable fraction was obtained by shaking 10 g of soil sample with 100 ml of 2 M HNO₃ for 1 h. The content of metals was measured in the filtered extracts by inductively coupled plasma-atomic emission spectroscopy (Spectro Analytical Instruments).

In order to determine the heavy metal concentrations in the leaves of bilberry, plant material was washed in a tap with distilled water and dried at 105 °C. A 0.25-g portion of dried plant material was treated with 5 ml of concentrated nitric acid and left for 24 h. Next, the samples were digested at 110 °C until complete mineralization was achieved. After mineralization, the samples were diluted with deionized water to a volume of 10 ml. Concentration of Cd, Pb, Zn, Fe and Mn was measured using inductively coupled plasma-atomic emission spectroscopy (Spectro Analytical Instruments). The quality of the analytical procedure was checked using a reference material (certified reference material CTA-OTL-1 Oriental Tobacco Leaves) with the same quantities of samples.

Analysis of the biochemical parameters of the plants

Protein content was determined according to the method of Bradford (1976) using bovine serum albumin as a standard. Proline content was determined by the method of Bates et al. (1973). The plant material (0.5 g) was homogenized in 10 ml of sulfosalicylic acid (3 g per 100 ml), and the homogenate was filtered through Whatman No. 2 filter paper. The reaction mixture containing 2 ml of homogenate, 2 ml acid ninhydrin and 2 ml of glacial acetic acid was incubated at 100 °C for 1 h. The reaction mixture was placed on ice and extracted with 4 ml of toluene. The absorbance was read at 520 nm using toluene as the blank. The proline content was expressed in micromoles proline per gram of fresh weight.

To measure the GSHt concentration, plant parts (0.5 g) were homogenized in trichloroacetic acid (5 g per 100 ml) and 0.125 mM phosphate buffer (pH 6.3) with 6.3 mM ethylenediaminetetraacetic acid (EDTA) and were centrifuged at 10,000×g for 10 min at 4 °C. Supernatants were used for GSH determination using the DTNB–GSSG reductase recycling procedure according to Anderson (1985). The reaction mixture contained 0.2 ml of supernatant, 0.6 ml of 0.3 mM NADPH, 0.1 ml of 6 mM DTNB and 0.1 ml (0.5 IU ml⁻¹) of glutathione reductase. The linear changes in the absorbance of the reaction mixtures were measured at 412 nm, and the GSHt was expressed as micromoles GSH per gram fresh weight.

The content of non-protein thiols was estimated as described by Mass et al. (1987). The plant material was homogenized in a 5 vol/g mixture containing 5-sulphosalicylic acid (2 g per 100 ml) and 1 mM EDTA and sodium ascorbate (0.15 g per 100 ml). The samples were centrifuged at 20,000×g for 10 min at 4 °C. Then, a 0.5-ml liquid supernatant, 0.5 ml of a 1 M sodium phosphate buffer (pH 8.0) and 100 μl of 10 mM 5,5′-dithio-bis (2-nitrobenzoic acid) (DTNB) were put into test tubes. The absorbance at 415 nm was read 1 min after the addition of DTNB. The number of non-protein SH groups was established based on a curve prepared using l-cysteine and expressed as nanomoles –SH per gram fresh weight.

For the analysis of guaiacol peroxidase, fresh plant material was homogenized in a 100-mM phosphate buffer (pH 6.8). The guaiacol peroxidase (GPX) activity was measured at 470 nm according to Fang and Kao (2000) and Liu et al. (2004) with guaiacol as the substrate. The GPX activity was measured in a reaction mixture (3 ml) containing a 50 mM phosphate buffer (pH 5.8), 1.6 μl H₂O₂, 1.5 μl guaiacol and 0.2 ml enzyme extract. The activity was calculated using the extinction coefficient (26 mM⁻¹ cm⁻¹) for tetraguaiacol and was expressed in micromoles tetraguaiacol per gram fresh weight per minute.

Statistical analysis

The data concerning the biochemical parameters and metal content were analysed, checked for normality and equality of variance. The data were analysed by ANOVA, and the treatments were treated as the independent variables. Significant statistical differences of all variables were established using the Tukey’s test (ANOVA; Statistica 10 package). We also calculated the Pearson’s correlation coefficient between the metal concentrations in separate soil extracants and in the leaves of *V. myrtillus*, and between the metal concentrations and biochemical parameters in the bilberry leaves. CANOCO 4.5 was used to carry out principal component analysis (PCA) which assessed the similarities and relations between heavy metal concentrations and biochemical parameters in the leaves of *V. myrtillus* in the studied areas.

Results

Heavy metal content and their availability in soil

There were significant differences in the content of the metals studied (HNO₃ extracted and CaCl₂ extracted)
between the polluted and potentially clean sites (Tables 2 and 3). Additionally, there was a clear difference in the concentrations of metals between the fraction of soil extracted with HNO₃ and the fraction extracted with CaCl₂. Among the metals examined, the highest concentrations of Fe, Pb and Zn were measured in the acid-extracted soil fraction. A several times lower concentration of the metals examined was determined in the fraction of soil extracted with CaCl₂. The following descending order of bioavailability was found among the heavy metals: Mn>Zn>Cd>Pb>Fe.

Heavy metal concentration in plants

The mean values of heavy metal concentrations in *V. myrtillus* leaves were found in descending order: Mn>Fe>Zn>Pb>Cd. There was a clear increase in the concentration of the metals studied in leaves of *V. myrtillus* at polluted sites. The exception was Mn, with the highest content observed at P4, PB5 and K6 sites. In most cases, metal content in *V. myrtillus* leaves increased with each month of sampling (Table 4). We found a strong positive correlation between concentrations of Cd, Pb and Zn in separate soil extractants and in leaves of bilberry. The range of coefficients was 0.61–0.89 with *p*<0.05. There was no correlation between the Fe and Mn concentrations in leaves and Fe and Mn concentrations measured in either of the two extractants (Table 5).

The biochemical status of the plants

At most sites, the total content of glutathione was by far the highest in May. The highest accumulation *GSHt* was found in the leaves of *V. myrtillus* from the M1 site—236.84 μmol GSHt g⁻¹ fresh weight (Fig. 2). There were no clear differences in glutathione content between contaminated and potentially non-contaminated sites. In addition, there was a negative correlation between the GSHt content and the concentration of Fe and Mn in the leaves of *V. myrtillus* (Table 6).

The greatest concentration of non-protein–SH groups in the leaves of *V. myrtillus* during the entire growing season was observed in the most contaminated M1 site (1,497.33–1,800 nmol –SH g⁻¹ fresh weight) (Fig. 3). This dependence was confirmed by a positive correlation between non-protein–SH groups and the content of Cd, Pb and Zn in the leaves of bilberry (Table 6). At the other sites, the content of non-protein thiols (NPTs) was the lowest in May and the highest in July. As with the NPTs, the highest protein content was also noted at the M1 site (4.28 mg g⁻¹ fresh weight; Fig. 4). A significant positive correlation between protein content and concentration of Cd, Pb and Zn was found in the leaves of *V. myrtillus* (Table 6).

Guaiacol peroxidase activity was by far the highest in the leaves of *V. myrtillus* from the L2 site (5.43–9.60 μmol tetraguaiacol g⁻¹ fresh weight min⁻¹), where the leaves of bilberry had the highest concentration of iron (Fig. 5). Meanwhile, the lowest GPX activity was observed in July and September at M1 (0.63–0.67 μmol tetra-guaiacol g⁻¹ fresh weight min⁻¹; Fig. 5). Guaiacol peroxidase activity in leaves of bilberries was positively correlated with Fe content, and negatively with the concentration of Cd, Pb and Zn (Table 6).

There were no significant differences in proline content between the contaminated and potentially non-contaminated sites (Fig. 6). There were also no significant correlations between proline content and the concentration of the examined metals in leaves of *V. myrtillus* (Table 6). PCA analysis

| Table 2 | The concentration of selected metals (milligrams per kilogram) in fractions of the soils extracted with HNO₃ (mean values±SE, n=5) |
|---------|--------------------------------------------------|
| Element | Month site | M1 | L2 | J3 | P4 | PB5 | K6 |
| Cd      | May       | 19.71±2.35 c | 7.80±0.61 b | 8.79±0.16 b | 2.76±0.26 a | 2.94±0.06 a | 0.68±0.00 a |
|         | July      | 13.72±0.17 e | 4.76±0.20 d | 4.23±0.02 c | 3.07±0.05 e | 4.77±0.03 d | 0.45±0.02 a |
|         | September | 25.54±0.56 d | 4.22±0.20 b | 7.86±0.70 c | 1.33±0.06 a | 5.30±0.37 b | 0.85±0.03 a |
| Fe      | May       | 931.40±65.50 a | 2,321.90±107.00 d | 2,960.90±82.00 e | 1,205.40±103.50 b | 3,428.40±2.50 f | 1,577.40±33.50 c |
|         | July      | 1,151.40±11.00 a | 1,345.90±55.00 b | 1,947.90±25.00 e | 1,208.40±11.50 a | 1,652.40±6.50 d | 1,545.90±6.00 c |
|         | September | 1,460.40±59.50 b | 1,247.90±9.00 a | 3,586.40±108.50 e | 2,129.90±44.00 c | 2,658.90±22.00 d | 2,560.40±10.50 d |
| Mn      | May       | 35.27±2.95 b | 194.30±13.90 d | 111.55±1.35 c | 17.20±0.57 a | 8.59±0.28 a | 6.84±0.33 a |
|         | July      | 49.96±0.54 d | 94.70±4.70 e | 47.41±0.86 c | 26.70±0.22 b | 7.53±0.06 a | 3.73±0.09 a |
|         | September | 39.42±2.21 b | 71.35±1.35 c | 120.85±7.95 d | 1.89±0.23 a | 8.48±0.64 a | 4.43±0.22 a |
| Pb      | May       | 674.49±76.50 c | 362.14±20.15 c | 481.14±3.15 d | 224.54±19.85 ab | 316.44±7.95 bc | 175.84±40.25 a |
|         | July      | 1,655.49±8.50 c | 239.44±10.15 c | 287.89±2.00 d | 253.64±3.05 c | 182.59±0.50 b | 81.64±1.25 a |
|         | September | 1,874.99±45.00 c | 235.19±6.90 bc | 524.44±35.55 d | 174.74±8.15 ab | 265.79±8.90 c | 124.59±0.90 a |
| Zn      | May       | 856.50±97.50 d | 559.50±39.50 c | 486.85±11.35 c | 144.90±13.30 b | 85.60±0.70 ab | 19.17±0.43 a |
|         | July      | 430.75±3.55 f | 321.40±10.20 e | 234.20±0.80 d | 149.80±2.20 c | 76.80±0.10 b | 36.80±8.37 a |
|         | September | 871.50±2.50 c | 260.65±12.35 c | 479.80±33.20 d | 47.69±4.01 a | 149.45±8.45 b | 54.55±3.05 b |

The different letters denote significant differences between the particular metal concentrations in the same month (*p*<0.05)
confirmed a strong correlation between the concentration of Cd, Pb and Zn, and the content of non-protein – SH groups and protein. It also showed a strong link between NPTs and total glutathione content. These relationships are best reflected in the most polluted area (M1). The analysis also demonstrated the close correlation between the activity of GPX and Fe content in leaves of *V. myrtillus* at the L2 site (Fig. 7).

### Discussion

The prediction of bioavailability of metals is of crucial importance for the assessment of environmental quality of contaminated soil (Gupta and Sinha 2007). The distribution and abundance of total metal concentrations are useful indicators of the extent of soil contamination (Tokalioglu et al. 2000; Xia et al. 2011; Dao et al. 2012), but risk from

### Table 3

| Element | Month | Site | L2 | J3 | P4 | PB5 | K6 |
|---------|-------|------|----|----|----|-----|----|
| Cd      | May   | 5.44±0.19 d | 0.36±0.00 ab | 0.83±0.12 c | 0.79±0.14 bc | 0.82±0.10 c | 0.18±0.04 a |
|         | July  | 4.82±0.15 d | 0.20±0.00 a  | 0.51±0.01 b  | 0.46±0.01 b   | 1.34±0.01 c | 0.16±0.01 a |
|         | September | 8.59±0.34 d | 0.50±0.02 ab | 0.87±0.01 b  | 0.58±0.01 ab  | 1.31±0.08 c | 0.29±0.00 a |
| Fe      | May   | 1.79±0.01 a  | 1.66±0.07 a   | 14.51±11.34 ab | 26.48±0.42 b   | 61.90±7.45 c | 24.31±5.05 b |
|         | July  | 1.84±0.02 b  | 1.18±0.02 a   | 3.60±0.17 c   | 4.31±0.07 d   | 22.62±0.24 e | 35.68±0.32 f |
|         | September | 14.25±10.44 b | 1.02±0.01 a   | 6.91±0.42 ab  | 41.30±2.45 c   | 46.80±1.25 c | 48.10±0.75 c |
| Mn      | May   | 10.82±0.65 a | 10.65±0.38 a  | 12.68±6.79 a  | 7.91±0.13 a   | 9.67±1.49 a  | 6.24±1.37 a  |
|         | July  | 14.17±0.87 b | 3.54±0.02 a   | 4.25±0.11 a   | 3.94±0.03 a   | 12.86±7.52 b | 2.74±0.11 a  |
|         | September | 24.54±0.32 c | 6.99±0.31 b   | 24.89±11.18 c | 2.32±0.01 a   | 6.69±0.09 a  | 3.06±0.52 a  |
| Pb      | May   | 3.36±2.61 a  | 1.14±0.11 a   | 6.74±2.19 b   | 4.53±0.07 a   | 2.72±0.01 a  | 1.75±0.74 a  |
|         | July  | 37.20±0.75 d | 0.96±0.06 a   | 8.97±0.10 c   | 2.63±0.27 a   | 2.82±0.08 b  | 1.56±0.18 a  |
|         | September | 22.29±1.80 c | 1.24±0.49 a   | 3.43±0.54 a   | 8.46±0.10 b   | 3.50±0.04 a  | 1.86±0.64 a  |
| Zn      | May   | 341.00±13.50 c | 37.18±0.57 a  | 4.65±0.02 a   | 97.93±1.98 d  | 73.93±0.25 f | 20.75±0.10 a |
|         | July  | 212.28±8.52 d | 21.60±0.44 b  | 39.70±0.30 c  | 36.85±0.80 c  | 32.50±0.25 c | 8.57±0.08 a  |
|         | September | 477.75±0.25 f | 49.00±2.50 c  | 83.35±0.45 e  | 30.35±0.35 b  | 73.93±1.98 d | 20.75±0.10 a |

The different letters denote significant differences between the particular metal concentrations in the same month (*p*<0.05)

### Table 4

| Element | Site | Month | L2 | J3 | P4 | PB5 | K6 | Sufficient or normal* | Excessive or toxic* |
|---------|------|-------|----|----|----|-----|----|----------------------|--------------------|
| Cd      | May  | 1.21±0.13 c | 0.86±0.09 b | nd  | 0.34±0.02 a | nd  | nd  | 0.05–0.2             | 5–30               |
|         | July | 1.63±0.11 c | 0.86±0.09 b | nd  | 0.41±0.01 a | nd  | nd  | 0.09±0.07 a          |                   |
|         | September | 6.26±0.22 d | 0.88±0.07 c | nd  | 0.44±0.01 b | 0.98±0.14 c | 11.26±2.21 a |                   |                   |
| Fe      | May  | 68.00±0.48 e | 127.75±2.00 f | 53.45±0.83 d | 50.54±0.52 c | 14.73±0.07 b | 9.36±0.00 a |                   |                   |
|         | July | 68.97±0.00 e | 164.29±0.00 e | 52.95±2.70 b | 13.52±9.50 a | 90.25±8.67 d | 71.74±2.38 c |                   |                   |
|         | September | 118.24±1.15 d | 409.49±6.33 f | 50.94±0.65 b | 90.17±16.48 c | 240.15±9.00 e | 1,138.99±103.67 e |                   |                   |
| Mn      | May  | 38.82±1.03 a | 74.81±2.52 b | 182.49±1.50 c | 282.99±9.90 e | 267.16±0.17 d | 257.66±1.00 d |                   |                   |
|         | July | 45.99±1.93 a | 97.84±0.68 a | 280.82±28.50 b | 460.49±71.83 c | 1,138.99±103.67 e | 917.66±1.33 | 30–300               | 400–1,000          |
| Pb      | May  | 45.99±3.77 a | 99.47±3.82 a | 397.99±1.33 b | 600.99±14.33 c | 1,056.16±69.83 d | 610.82±145.50 c |                   |                   |
|         | July | 35.88±6.08 b | 13.44±1.43 a | 10.97±6.96 a  | 3.56±0.35 a   | 9.39±0.94 a   | 9.39±2.42 a   |                   |                   |
|         | September | 54.94±3.32 c | 7.56±0.59 a  | 8.77±0.81 a   | 10.40±3.09 a  | 23.83±3.70 b  | 14.39±2.58 a  | 5–10                 | 30–300             |
| Zn      | May  | 69.73±0.63 f | 19.82±0.45 d | 13.48±0.48 b  | 3.55±0.13 b   | 38.60±2.40 d  | 17.08±0.15 a  | 27–150               | 100–400            |
|         | July | 97.98±0.08 e | 26.38±0.18 c | 21.58±1.18 b  | 35.53±0.13 b  | 38.60±2.40 d  | 17.08±0.15 a  | 27–150               | 100–400            |
|         | September | 207.17±5.50 d | 36.80±0.13 b | 17.90±0.27 a  | 21.12±1.75 a  | 72.32±0.25 c  | 20.78±4.29 a  |                   |                   |

The different letters denote significant differences between the particular metal concentrations in the same month (*p*<0.05)

* According to Kabata-Pendias and Pendias 2001
metals depends on their bioavailability (Prokop et al. 2003; van Gestel 2008; Dao et al. 2012). A large part of biomonitoring research consists in the comparison of bioavailable fractions of metals (extracted with CaCl2) with their total concentration or pseudo-total (extracted, e.g. 2 M HNO3) in soil (Ullrich et al. 1999; Peijenburg and Jager 2003; Pueyo et al. 2004; Keller and Hammer 2004). The assessment of the environmental risk requires the determination of not only total metal concentrations in soils but also the bioavailable fraction (Dao et al. 2012).

There was a clear difference between the concentration of the studied metals in the potentially bioavailable extracted fraction of CaCl2, and HNO3 extracted fraction at all the studied sites (Tables 2 and 3). Some differences in the metal concentration sequences in the examined soil extracts and leaves of V. myrtillus were found. The order of metals in the soil fraction extracted with HNO3 was as follows: Fe>Pb>Zn>Mn>Cd, and for the CaCl2 extracts, it was Mn>Zn>Cd>Pb>Fe, while the range of metal contamination in the leaves was in the following order: Mn>Fe>Zn>Pb>Cd.

A similar trend was observed by Celik et al. (2005) in Robinia pseudoacacia (Fe>Mn>Zn>Pb>Cu>Cd), Kafel et al. (2010) in the leaves of Philadelphus coronarius (Zn>Ni>Pb>Cu>Cd) and Kozanecka et al. (2002) in different plant species from the non-polluted areas of eastern Poland (Zn>Cu>Ni>Pb>Cr>Cd). An increase of Cd, Pb and Zn in the leaves of V. myrtillus was strongly associated with the increased concentrations of these metals in both studied soil fractions (Table 5). A significantly higher level of Cd, Pb and Zn was found in the leaves of V. myrtillus collected at the site located near the zinc smelter (M1) in comparison with other potentially contaminated areas and those free of contamination (Table 4). Levels of these three metals exceeded values considered as normal, and at the end of the growing season, in September, the levels exceeded values considered toxic for plants (according to Kabata-Pendias and Pendias 2001). Higher content of Zn (298 mg kg$^{-1}$), but lower Pb (8.93 mg kg$^{-1}$) and Cd (3.20 mg kg$^{-1}$) than in the presented study were found in the leaves of bilberries by Bialońska et al. (2007) near the Zn–Pb smelter in Bukowno.

The concentration of Mn (Table 4) in the leaves of V. myrtillus at sites classified as unpolluted (P4, PB5 and K6) in most cases exceeded values considered as toxic to plants (according to Kabata-Pendias and Pendias 2001). The ericaceous species, especially V. myrtillus, are distinguished by a high element content of Mn, irrespective of the content in the soil (Reimann et al. 2001; Salemaa et al. 2004; Mróz and Demczuk 2010). Mróz and Demczuk (2010) suggested that V. myrtillus is an accumulator of Mn, and such high concentrations of this element suggest a possibility that bilberry uses Mn for some beneficial purposes. At the other sites, the concentration of Cd, Pb, Zn, Fe and Mn was similar to other uncontaminated sites reported in the literature (Reimann et al. 2001; Kozanecka et

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**Table 5** The correlation coefficients between the concentrations of particular metals in separate soil extracants (with HNO3 or CaCl2) and in the leaves of *V. myrtillus* ($p<0.05$)

|          | Cd    | Fe    | Mn    | Pb    | Zn    |
|----------|-------|-------|-------|-------|-------|
| HNO3     | 0.81  | 0.68  | NS    | NS    | NS    |
| CaCl2    | 0.68  | NS    | NS    | NS    | 0.85  |

**Notes:**
- NS not significant
The decline in the glutathione content (Rao and Sresty 2000). Sudhakar et al. (2006) showed that to play some role in the synthesis of phytochelatins (Madhava Rao and Sresty 2000; Bat Mohamed 2010; Bat Platanus orientalis; Vicia faba—El-Beltagi and Schützendübel 2001; Pb at Raphanus sativus—El-Beltagi and Mohamed 2010; Pb at Vicia faba and Phaseolus vulgaris—Piechalak et al. 2002). The decline in the glutathione content of plants may result from the inhibition of the enzymes involved in glutathione synthesis by toxic metal ions. In addition, the depletion of glutathione pool may also be considered to play some role in the synthesis of phytochelatins (Madhava Rao and Sresty 2000). Sudhakar et al. (2006) showed that exposure of Hydrilla verticillata (L.f) Royle to high doses of copper led to a decrease in GSH and an elevation in PC levels. Boojar and Tavakoli (2010) found a similar dependency in Alhagi camemelorum Fisch.

Tolerance toward metals correlates well with the level of non-protein thiols which include not only glutathione but also thiol-rich peptides known as phytochelatins and other SH-rich compounds (Kafel et al. 2010). Molecules containing sulphur, which exist in a wide variety in cells, may fulfill different functions and may be independently regulated (Mishra et al. 2009). In our study, an increase of the NPT content was noticed in the leaves of V. myrtillus grown at a polluted site (Fig. 3). Additionally, the content of non-protein thiols was positively related with concentrations of Cd, Pb and Zn (Table 6 and Fig. 7). This was similar to the study of Nadgórska-Socha et al. (2011) who observed an increase in non-protein –SH group content in leaves of Silene vulgaris populations on the substrate with Cd and combination of metals (Zn, Cd and Pb). Also, Mishra et al. (2009) found a strong positive correlation between NPTs and Cd content in plants of Ceratophyllum demersum L. The increase in non-protein thiols levels indicates an ability to tolerate the cellular metal load (Mishra et al. 2006).

One of the mechanisms affected by heavy metals in plants is protein synthesis. It is known that soluble protein content is an important indicator of physiological status of plants (Doganlar and Atmaca 2011). In this study, we noticed a significant increase in protein content of bilberry leaves from most polluted site (Fig. 4), and we found a strong positive correlation between protein content and Cd, Pb and Zn concentrations (Table 6 and Fig. 7). Similarly, Doganlar and Atmaca (2011) reported that an increase in total soluble protein contents related to Al concentration in leaves of Platanus orientalis and Zn concentration in leaves of Melia azaderach. Mesmar and Jaber (1991) found increased total protein with increasing lead concentration in wheat and lentil, and Singh et al. (2006) reported increased protein contents in Oryza sativa plants under Cd stress. The increase in protein

| Cd  | Fe  | Mn  | Pb  | Zn  |
|-----|-----|-----|-----|-----|
| GSHt| NS  | −0.30| −0.33| NS  | NS  |
| NPTs| 0.31| NS  | −0.38| 0.38| 0.40|
| Protein| 0.66| NS  | NS  | 0.64| 0.65|
| GPX | −0.29| 0.30| NS  | −0.36| −0.37|
| Proline| NS  | NS  | NS  | NS  | NS  |

NS not significant

Fig. 3 Non-protein –SH groups content (nanomoles –SH per gram fresh weight) in V. myrtillus leaves (mean values±SE, n=5). Different letters above the columns indicate significant differences in the same month (p<0.05)
content may be a specific mechanism thanks to which cells compensate for the protein content that have been deactivated due to metal binding, and also the effect of the increasing the content of stress proteins (Mesmar and Jaber 1991; Seregin and Ivanov 2001).

Peroxidases are antioxidant enzymes which play a crucial role in plant growth and development, and activities of these enzymes are charged under both abiotic and abiotic stress conditions (Doganlar and Atmaca 2011). In this study, in the leaves of *V. myrtillus* the highest GPX activity was observed in the vicinity of iron smelter at the L2 site. The lowest values of guaiacol peroxidase activity were at M1 (Fig. 5). An increase in GPX activity in leaves of *V. myrtillus* was observed under the influence of Fe and its reduction in the presence of Cd, Pb and Zn (Table 6 and Fig. 7). Many authors reported increased GPX activity increased content of heavy metals (Macfarlane and Burchett 2001; Markkola et al. 2002; Hagemeyer 2004; Kafel et al. 2010; El-Beltagi and Mohamed 2010; Doganlar and Atmaca 2011). The positive relationship between metal content and GPX activity was also reported by Nadgórska-Socha et al. (2008) in the leaves of *Philadelphus coronarius* Linne from a highly urbanized area of Krakow. However, Baycu et al. (2006), in their examination of peroxidase activity in leaves of *Acer* and *Alianthus* growing in the urban parks of Turkey, observed both increased and decreased peroxidase activity compared to controls. Pongrac et al. (2009) reported no change in GPX activity in *Thlaspi praecox* and *Thlaspi caerulescens* in the presence of Cd and Zn; similar results were obtained by Gratão et al. (2008) in leaves, roots and fruits of tomato grown in conditions of cadmium contamination. Similar to our work, Sandalio et al. (2001) observed a decrease in peroxidase activity and other antioxidant enzymes in pea under the influence of Cd. Similar to the work of Stobrawa and Lorenc-Plucińska (2007), at most sites GPX activity was the highest in spring. Perhaps this is connected with the fact that peroxidases, while controlling the level of hydrogen peroxide, at the same time catalyze the synthesis of lignin–phenol polymers, a process that is most intense after the winter rest, i.e. in spring (Stobrawa and Lorenc-Plucińska 2007).

Accumulation of free proline in response to the impact of heavy metals is a widespread phenomenon among many plant species (Schat et al. 1997, Chen et al. 2004, Tripathi...
and Gaur 2004; Balestrasse et al. 2005; Sharma and Dietz 2006; Abdel-Latif 2008). Numerous studies have shown that the content of free proline depends on plant species and varies between organs. Moreover, proline concentration depends on the type of metal and its concentration (Sharma and Dietz 2006; Abdel-Latif 2008). Nikolić et al. (2008) found a significantly higher accumulation of free proline in the roots than in leaves of hybrid poplar under the influence of cadmium. Sharma and Dietz (2006) reported an increase in free proline accumulation in response to increased accumulation of heavy metals in several plant species (Cajanus cajan, Vigna mung, Helianthus annus, Lemma minor, Triticum aestivum, Lactuca sativa, Silene vulgaris, O. sativa). However, our tests have not found any significant effect of heavy metals on the accumulation of proline in the leaves of V. myrtillus. Proline content in leaves of bilberries only slightly differed between the various research sites. It is difficult to draw unambiguous, major conclusions from research carried out in the field conditions because such experiments can be affected by environmental factors.
Nevertheless, multi-parameters approach provided understanding of the diverse responses and effects of exposure to contaminants, and the effective risk it poses for different plant species (Stobrawa and Lorenc-Plucińska 2007; Fonseca et al. 2011; Oliva et al. 2012).

Based on our study, we propose the following conclusions: An increase in content of protein and non-protein SH groups and changes in GPX activity in the leaves of V. myrtillus from polluted areas could be the evidence of an enhancement of the oxidative stress. Since these parameters seem to be universal, sensitive and correlated well with heavy metals stress, they could be good biochemical tools to predict heavy metal pollution.

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