**Acacia longifolia**: A Host of Many Guests Even after Fire

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**Abstract:** *Acacia longifolia* is a worldwide invader that cause damage in ecosystems, expanding largely after wildfires, which promote germination of a massive seed bank. As a legume, symbiosis is determinant for adaptation. Our study aims to isolate a wider consortium of bacteria harboured in nodules, including both nitrogen and non-nitrogen fixers. Furthermore, we aim to evaluate the effects of fire in nodulation and bacterial diversity on young acacias growing in unburnt and burnt zones, one year after the fire. For this, we used molecular approaches, M13 fingerprinting and 16S rRNA partial sequencing, to identify species/genera involved and δ^{15}N isotopic composition in leaves and plant nodules. Nitrogen isotopic analyses in leaves suggest that in unburnt zones, nitrogen fixation contributes more to plant nitrogen content. Overall, *A. longifolia* seems to be promiscuous and despite *Bradyrhizobium* spp. dominance, *Paraburkholderia* spp. followed by *Pseudomonas* spp. was also found. Several species not previously reported as nitrogen-fixers were identified, proposing other functions besides ammonia acquisition. Our study shows that bacterial communities are different in nodules after fire. Fire seems to potentiate nodulation and drives symbiosis towards nitrogen-fixers. Taken together, a multifunctional community inside nodules is pointed out which potentiate *A. longifolia* invasiveness and adaptation.

**Keywords:** climate change; invasive species; nodulation; symbiosis; *Bradyrhizobium* spp.

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**1. Introduction**

In a fast-changing planet and under a climate change scenario, biological invasions have become a serious problem. To overcome them, understanding species mechanisms to adapt to a new environment is crucial for biodiversity protection [1], conservation ecology and management strategies [2]. Exotic species introduction and outcompeting with natives can lead to invasion [3], due to their ability to easily adapt to new environments. One of the largest and widespread families of flowering plants is Fabaceae, which includes species that are becoming major threats for biodiversity [4].

*Acaea* is one of such genera and it constitutes a polyphyletic group comprising over 1350 species [5], the majority native to Australia [6]. Several acacias have been introduced outside Australia and have resulted in invasive populations worldwide, present at a higher frequency in Mediterranean climates like California, the Iberian Peninsula, and South Africa [7]. *Acacia longifolia* (Andrews) Willd. (Sidney golden wattle) is considered nowadays one of the most aggressive species worldwide as well as one of the most interesting invaders. In Portugal, this species was introduced for dune stabilization, preservation of sand erosion and with ornamental purposes during the late nineteenth century and the
beginning of the twentieth [8]; however, nowadays it presents an ecological concern. In fact, *Acacia* spp. contributes largely to the 18% of the Portuguese mainland area occupied by alien species [9]. *A. longifolia* has been described as an “ecosystem-engineer” [10–12], due to its specific characteristics, including: high adaptive plasticity, altering soil environment, ecosystems’ functioning and diminishing local biodiversity [1], formation of a soil seed bank with a high germination rate and seedlings survival [13] and high water and resources consumption [14,15]. The control measures of this species are challenging and should be added as a priority in applied conservation research.

Moreover, this “engineering” is mostly determined by its inherent ability to fix atmospheric nitrogen through Legume–Rhizobia symbioses. Biological nitrogen fixation occurs inside nodules, developed in roots, which includes a complex and reciprocal signalization process between the host plant and compatible bacteria. The host plant releases flavonoids that stimulate the production of Nod factors by bacteria and triggers root invasion and nodule formation. Besides this structural development, the presence of nitrogenase complex (*nif* genes) is essential, allowing nitrogen fixation, and leghemoglobin responsible for microaerophilic environments as an oxygen-buffer, for nitrogenase functioning inside nodules. When everything is perfectly settled, bacteria provides usable nitrogen forms to the plant in exchange for carbohydrates [16].

Invasiveness, and its ecosystems impacts, is in fact a process highly mediated by plant–microbe interactions [17], with several gaps remaining in our knowledge on this process. *Acacia* spp. does not have a specific bacterial partner, and the identification of who is taking part in nodule’s bacteriome remains a major challenge. Until now, several studies showed *Bradyrhizobium* as the most common partner in native and introduced environments, followed by *Rhizobium* [18], while *Mesorhizobium* and *Sinorhizobium* appear mostly in the native range [19–22]. In fact, nitrogen-fixers are among the functional and taxonomically diverse rhizosphere communities [17]. *A. longifolia* is a promiscuous woody species, that possibly exploits soil bacterial diversity to find partners for symbiotic success, particularly relevant in the context of invasion [23,24]. Studies have been focused so far on nitrogen-fixing species, with a lack of information on other partners identification.

Under climate change, fire events are becoming more frequent and are an important and crucial area to address. Fires are a common disturbance in regions with Mediterranean climates (such as Portugal and Australia) [25]. Since fire stimulates acacia seed germination [16], a burst in natural regeneration ultimately results in *Acacia* sp. dominance. Despite the importance of understanding young plants’ invasive behaviour, particularly concerning the establishment of symbiotic interactions, no studies have been reported so far. Finding the adequate symbiotic partners can determine the success or the decline of the invader, highlighting the importance of clarifying the belowground relation between microbial communities and the invader *A. longifolia*. Furthermore, nodules can harbour more than nitrogen-fixing rhizobia, as has been shown by Martinez-Hidalgo and Hirsch [26], and the total diversity needs to be studied.

In this paper, (1) we aim to address the wider consortium of bacteria that nodules harbour, including bacterial communities not involved in nitrogen fixation. Additionally, (2) we hypothesise that fire could affect nodulation per se, rendering after fire root nodules’ bacterial community different. With that in mind, understanding above- and belowground dynamics is a key factor to understand *A. longifolia* as an ecosystem transformer.

2. Materials and Methods

2.1. Study Site Description and Nodules Collection

Nodules from young *A. longifolia* plants were collected in Mira, Aveiro, Portugal mainland (40.52451° N, 8.67253° W), one year after fire occurrence, in October of 2017. This region has a Mediterranean climate, with an Atlantic influence. In the last 30 years, annual temperatures ranged from a mean minimum of 10 °C to a maximum of 20.2 °C and a mean annual precipitation of 904 mm. However, 2017 was a harsh year, registering a hot and dry Spring with the hottest April since 1931 and
almost no precipitation registered, leading to an increasing of drought from May until beginning of the Autumn, culminating in the October fires [27].

The study site was an area occupied with forest, including *Acacia* sp., *Eucalyptus* sp. and *Pinus* sp. trees. Six sampling sites were selected, including three unburnt zones or zones were no fire occurred (UBZ) and three burnt zones (BZ) (Figure 1). An area of 25 m² (5 × 5 plots) was established in each sampled zone, where eight individual young plants (20–60 cm) were selected randomly. *Acacia* nodules were collected by digging up plants to identify roots with attached nodules. Nodules were counted, their size was measured and then they were stored in silica gel and kept under room temperature until use.

![Schematic representation of the location of the six sampled sites: three unburnt zones (green) and three burnt zones (red). The 5 × 5 m plots were established to collect young plants and nodules.](image)

2.2. Soil Characteristics

Soil samples were collected from a depth of 0–20 cm after removing litter layer. A mixed sample was made through the collection of soil from three spots in each zone according to Sankhla et al. [28], and each sample was approximately 1.5 kg. For soil analysis, three subsamples were collected and mixed in a composite sample; these samples were analysed for basic characteristics such as texture and particle size, pH (water and KCl 1 M) through a suspension method and potentiometry, organic matter (OM) with thermic decomposition, P₂O₅ (phosphorus) and K₂O (potassium oxide) and the amount of total and mineral N (N-NH₄⁺ and N-NO₃⁻), all through molecular absorption in a segmented flux analyzer. Soil was characterized by a coarse texture in both studied zones, with a pH of 5.4–5.5 (Table 1). Analysis was performed by Plants and Soils Laboratory from Universidade de Trás-os-Montes e Alto Douro (UTAD), Portugal.

| Texture | Water pH | OM (%) | P₂O₅ mg kg⁻¹ | Total N g kg⁻¹ | N-NO₃⁻ | N-NH₄⁺ |
|---------|----------|--------|--------------|----------------|--------|--------|
| UBZ     | 5.5      | 1.02   | 3            | 0.054*         | 2.8    | 2.8    |
| BZ      | 5.4      | 2.08   | 16           | 0.124          | 4.3    | 6.3    |

2.3. Isotopic Analysis

Leaves and nodules from young plants were collected from the six sampled sites and were dried during 48 h in a drying kiln at 60 °C. Each sample was ground using a ball mill and 2–2.5 mg was weighted for isotopic analysis. ¹³C/¹²C and ¹⁵N/¹⁴N ratios in the samples were determined using...
a continuous flow isotope mass spectrometry on a Sercon Hydra 20–22 (Sercon, Crewe, UK) stable isotope ratio mass spectrometer, coupled to a EuroEA (EuroVector, Pavia, Italy) elemental analyser for online sample preparation by Dumas-combustion. Delta (δ) calculation was performed according to δ = [(Rsample − Rstandard)/Rstandard] × 1000, where R is the ratio between the heavier and lighter isotopes. δ¹⁵N air values are referred to air and δ¹³C VPDB values are referred to PDB (Pee Dee Belemnite). The (secondary) reference materials used were Sorghum Flour Standard OAS/Isotope and Wheat Flour Standard OAS/Isotope (Elemental Microanalysis, UK) for nitrogen and carbon isotope ratio (with, respectively, δ¹⁵N air (Sorghum Flour OAS) = 1.58 ± 0.15‰, δ¹⁵N air(Wheat Flour OAS) = 2.85 ± 0.17‰, δ¹³C VPDB (Sorghum Flour OAS) = −13.68 ± 0.19‰, δ¹³C VPDB(Wheat Flour OAS) = −27.21 ± 0.13‰), regularly checked against certified reference materials. Uncertainty of the isotope ratio analysis, calculated using values from six to nine replicates of secondary isotopic reference material interspersed among samples in every batch analysis, was ≤0.1‰. The major mass signals of N and C were used to calculate total N and C abundances, using Sorghum and Wheat Flour Standard OAS (Elemental Microanalysis, UK, with 1.47% N, 46.26% C and 1.47% N, 39.53% C respectively) as elemental composition reference materials.

All the analyses were performed at the Stable Isotopes and Instrumental Analysis Facility (SIIAF), Faculty of Sciences of the University of Lisbon, Portugal.

2.4. Isolation and Phenotypic Characterization of Nodule Bacteria

For bacterial isolation, nodules were rehydrated in water during 12 h and surface-disinfected in 70% EtOH for 1 min, then transferred to commercial bleach for 6 min and 1 min in 70% EtOH, followed by six washes in sterile distilled water (adapted from Vincent [29]). For disinfection control, nodules were dried with sterile filter paper and rolled (surface printing) in YMA (Yeast Mannitol Agar) plates, incubated at 28 °C for four days. YMA was chosen, once it is selective to nitrogen-fixers, allowing pre-selection when isolation. Only in the absence of growth the nodules were processed, preventing the growth of microorganisms from the soil or nodule surface.

Pools of 1–4 nodules were crushed in 500 µL of 0.85% sodium chloride; by serial dilutions method (10⁰ to 10⁻³), each suspension was inoculated in YMA supplemented with 0.01% cycloheximide and incubated for 12 days. Pure cultures were obtained after three or more subculturing steps using standard protocols [30] (p. 16), [31].

Macroscopic appearance of bacterial growth was analysed, through patterns of growth described for nutrient agar plates (days of growth, size, pigmentation, form, margin, and elevation), according to Cappuccino and Sherman [30]. Routine tests, namely Gram staining and potassium hydroxide (KOH) test, catalase test and oxidase test were performed to cluster the colonies according to the results. Each of these tests has a dichotomic response, positive (+) or negative (-). All these tests were performed after 24 h of visible colonies growth.

2.5. DNA Fingerprinting of Bacterial Isolates

Genomic DNA was extracted using GES (Guanidium thiocyanate, Ethylenediamine tetracetic acid (EDTA) and Sarkosyl) modified protocol [32]. One loop of colonies from each isolate was used. For mucous colonies, five washes in sterile water were performed through suspension in 1 mL of autoclaved water before lysis. The protocol was then followed, and DNA was resuspended in 100 µL of 1× TE (Tris-EDTA).

Polymerase chain reaction (PCR) amplification for molecular fingerprinting was performed on a final volume of 25 µL containing 50 ng of template DNA, 1 U of Taq DNA polymerase (Invitrogen), 25 pmol of the primer csM13 (5’ GAGGGTGGCGGTTCT 3’; [33]), 3 mM of MgCl₂, 0.2 mM of each dNTPs and 1x PCR buffer. The PCR temperature profiles were 5 min followed by 40 cycles of 95 °C for 1 min, 50 °C for 2 min, 72 °C for 2 min and a final extension at 72 °C for 5 min. Amplification products were resolved through electrophoresis on 1% (w/v) molecular biology agarose gel dissolved in 0.5x TBE buffer. Gels ran at 85 V for 5 h. After running, they were stained in 0.5 µg·mL⁻¹ ethidium
bromide for 10 min, washed in water to remove excess staining and visualized under UV light in a transilluminator with the software Alliance 4.7 (Uvitec, Cambridge).

M13-PCR fingerprinting profiles were compared using BioNumerics software (Applied Maths, Sint-Martens-Latem, Belgium) and a dendrogram was performed using the Pearson correlation coefficient as association measure and the unweighted pair-group method with arithmetic mean (UPGMA) as clustering algorithm. Through this analysis, it was possible to cluster similar isolates, reducing our sample size, facilitating the selection to identification. An inherent limitation is ascribed to culture-dependent methods, which restricts our diversity analysis to cultivable microorganisms.

Reproducibility was analysed based on a random sample of 10% replicates of the total isolates, in order to establish the cut-off level. This represents the maximum level of similarity between two isolates to support their difference.

Shannon–Wiener and Simpson diversity indexes and Pielou evenness index [34] were used to calculate the diversity and evenness of the bacterial isolates of unburnt and burnt zones. A global approach was performed through cluster analysis, considering all the isolates obtained from unburnt zone and all from burnt zones. Indexes were then compared for further analysis.

2.6. Identification of Bacterial Isolates by 16S rRNA Gene Sequencing

After discriminating isolates by DNA fingerprinting, the 16S rRNA gene was amplified for a sub-set of isolates representative of almost all clusters, using two different primers combinations: PA(8f) with 907r or 104f with 1392r based on the *Escherichia coli* numbering system [35], depending on the success of amplification. SurfTaq (StabVida, Portugal) was the DNA polymerase used. The final volume of the PCR reaction was 20 µL, and the same master mix reagent concentration and temperature profiles were used, as previously described.

Following confirmation of a unique amplicon with the correct size, PCR products were purified using ExoSAP-IT™ PCR Product Cleanup Reagent (Thermo Fisher Scientific), according to the manufacturer’s protocol. After purification, the samples were sequenced through Sanger sequencing (StabVida, Portugal).

DNA sequences were analysed with the software Geneious [36], performing alignments among each other and later with data available from GenBank through BLAST. Each sequence was considered individually and aligned with sequences available in the GenBank. The alignments with high similarity were considered and taxonomic identification was achieved according to maximum pairwise identity. Through this identification, it was possible to putatively identify other non-sequenced isolates by comparing similarities through dendrogram analysis. Sequences were submitted to GenBank with accession numbers MT465339 to MT465388 (Supplementary Table S1).

2.7. Statistical Analysis

Principal component analysis (PCA) was performed with normalized values of the number of nodules, the soil properties and isotopic analysis, by subtracting the mean of each variable to each value and dividing by the standard deviation. This approach was used to explore differences between the 6 sampled zones in order to identify the main discriminatory variables. Mean number of nodules collected in all unburnt and burnt zones were analysed by t-test at a 95% confidence level (α = 0.05), as well as soil and isotopic data. All data collected were statistically analysed using packages FactoMiner and stats in R studio (v.3.6.1).

3. Results

3.1. Nodulation of Young Plants

In this study, 242 nodules were collected from unburnt zones and 337 nodules from burnt zones, for a total of 579 nodules. However, some variability was present in young acacias growing in sampled areas from both fire conditions. In unburnt zones 1, 2 and 3, 83, 128 and 31 nodules were counted,
respectively, while in burnt zones 1, 2 and 3 these values were 168, 95 and 74, respectively. Although no relevant differences in size and morphologies were detected between nodules from zones with and without fire (data not shown), a higher and more diverse nodulation index was found in burnt zones, with an average number of 14.0 nodules per young plant, against 10.1 nodules per young plant in unburnt zones. The differences found between treatments, however, were not statistically significant \((p > 0.05)\).

3.2. Isotopic Analysis

Nitrogen fixation efficiency in leaves was analysed through \(\delta^{15}N\), revealing values close to 0\% in both zones, in accordance with the occurrence of atmospheric nitrogen fixation through symbiosis. Despite no statistically significant differences between fire treatments, the values \(-1.0\%\) for unburnt and \(0.8\%\) for burnt ones were obtained from leaves, suggesting a higher nitrogen fixation in acacias growing in unburnt zones. In the nodules, no differences were found in \(\delta^{15}N\) values from unburnt and burnt zones \((7.9\%\) and \(7.4\%\) respectively). \(\delta^{15}N\) in the nodules remained extremely enriched (very positive values). Regarding \(\delta^{13}C\), there were no major differences between plants growing after fire conditions, both in leaves \((-30.2\) for unburnt and \(-29.5\) for burnt zones) and nodules \((-29.5\) for unburnt and \(-28.8\) for burnt zones) (see Table 2).

| Leaves |        |        |        | Nodules |        |        |        |
|--------|--------|--------|--------|---------|--------|--------|--------|
|        | \(\delta^{15}N\) | \(\delta^{13}C\) | %N | %C | C/N | \(\delta^{15}N\) | \(\delta^{13}C\) | %N | %C | C/N |
| UBZ    | -1.0   | -30.2  | 2.4   | 44.3    | 19.1   | 7.9    | -29.5  | 5.3 | 36.5 | 7.0 |
| BZ     | 0.8    | -29.5  | 3.3   | 43.5    | 15.2   | 7.4    | -28.8  | 6.3 | 42.3 | 7.0 |

PCA Analysis

PCA biplots showed that three main components explained 76.9\% of the variance. The unburnt and burnt zones are clearly separated in PC_1 (Dim1). This Dim1 explained 32.6\% of the total variance, mainly due to differences in \(\delta^{15}N_L\), \(P_2O_5\), \(K_2O\) and \(N-NO_3\). PC_2 (Dim2) explained 25.6\% of the variance, considering the OM, total_N and N-NH\(_4\), while PC_3 (Dim3) explained 18.7\% of the variance, mainly due to differences in the number of nodules. The bidimensional representation of both PC1_PC2 and PC1_PC3 (Figure 2a,b) reveals a clear separation between the burn and unburnt zones in both biplots, given the differences in the variables \(P_2O_5\), \(K_2O\), N-NO\(_3\), OM, total_N, N-NH\(_4\) and \(\delta^{15}N_L\). All the variables increased with fire, except \(\delta^{15}N_L\), which was lower in plants growing without fire, revealing a potentially higher nitrogen fixation. The difference in the number of nodules, higher after fire occurrence, is particularly relevant for PC3 (Figure 2b).

3.3. Bacterial Fingerprinting and Identification

A total of 153 isolates were obtained, 94 from the unburnt zone and 59 from the burnt zone. After phenotypic analysis, genomic fingerprinting based on csM13 was performed and the results are presented in the dendrogram of the Supplementary Figure S1. This dendrogram showed that isolates obtained from UBZ and BZ were scattered, and there were no clustering grouped accessions according to fire treatment. The isolates were mainly grouped according to its genera, although some genera with more representatives (e.g., \textit{Bradyrhizobium} sp.) were clustered in different groups. A total of 19 clusters were identified. Some of the isolates were clustered independently, forming a group with a unique representative.
Figure 2. Principal Component Analysis (PCA) of soil properties and isotopic analysis from all unburnt (UBZ) and burnt zones (BZ), represented by Dim1 and Dim2 in (a) and Dim1 and Dim3 in (b). Only the main contributory variables to Dim1 and Dim2 are represented in (a), which are $\delta^{15}$N_L ($\delta^{15}$N in leaves), P2O5 (phosphorus), K2O (potassium oxide), N_NO3 (nitrates) for Dim1 and OM (organic matter), Total_N (total nitrogen) and N_NH4 (ammonia) for Dim2; only the main contributory variables to Dim1 and Dim3 are represented in (b), which are the number of nodules for Dim3. Numbers 1, 2 and 3 represent the UBZ 1, UBZ 2 and UBZ 3 zones respectively, while 4, 5 and 6 represent the BZ 1, BZ 2 and BZ 3 zones.
16S rRNA gene sequencing allowed for the preliminarily identification of 50 isolates, with up to 94.2% pairwise identity (See Supplementary Table S1). *Bradyrhizobium* and *Paraburkholderia* were the most represented genera with 23 and 10 isolates, respectively, followed by *Pseudomonas*, represented by seven isolates. *Caballeronia*, *Duganella*, *Micrococcus*, *Moraxella*, *Paenibacillus*, and *Rhizobium* were also identified genera (Supplementary Table S1). These data were considered together with the dendrogram analysis, allowing the inference of the identification of other isolates belonging to the same cluster, following a previous similarity evaluation of the fingerprinting profile. As a result, more genera are represented in UBZ comparing to BZ. Considering the 153 isolates obtained, the 94 isolates from unburnt zone were distributed in five different classes: Alphaproteobacteria (39.4%), Betaproteobacteria (26.6%) and Gammaproteobacteria (16%) from phylum Proteobacteria; Actinobacteria (3.2%) from phylum Actinobacteria and Bacilli (2.1%) from phylum Firmicutes. Part of the collection remained unclassified, accounting for 12.8% of the isolates. The 59 isolates from burnt zone were distributed in four different classes: Alphaproteobacteria (45%), Betaproteobacteria (13.3%) and Gammaproteobacteria (10%) from phylum Proteobacteria; Actinobacteria (8.3%) from phylum Actinobacteria. Additionally, 23.3% of the isolates remained unclassified (Figure 3 and Table 3). A curious and unexpected result was the presence of only one cluster with three isolates identified as *Rhizobium* sp. and the absence of isolates from *Sinorhizobium* and *Mesorhizobium* genera. Regarding species identification, *Bradyrhizobium cytisi* is the most represented one (Supplementary Table S1). In fact, through the diversity and evenness indexes, we found a higher diversity in the unburnt zone and a dominance of a species, *Bradyrhizobium cytisi*, in both zones, as already mentioned. As shown by Shannon–Wiener diversity index, there was a higher diversity in isolates from the unburnt zone (H’ = 1.0) than from the burnt zone (H’ = 0.74). The Pielou evenness index revealed that in both zones some isolates were dominant (J’ = 0.75 for unburnt zone and 0.67 for burnt zone). The Simpson diversity index also showed a high diversity for both the unburnt zone (D’ = 0.97) and burnt zone (D’ = 0.98).

**Figure 3.** Percentage of the distribution of the different phyla/class of bacterial isolates from *A. longifolia* young plants nodules, present in both unburnt (UBZ) and burnt zones (BZ).
Table 3. Total of isolates identified within each unburnt (UBZ 1, UBZ 2, UBZ3) and burnt (BZ 1, BZ 2, BZ 3) zone. Colours indicate the phylum of each genus (Proteobacteria/α-proteobacteria (blue), Proteobacteria/β-proteobacteria (orange), Proteobacteria/γ-proteobacteria (green), firmicutes/Bacilli (purple) and actinobacteria/Actinobacteria (yellow).

| Genera/Zones       | UBZ 1 | UBZ 2 | UBZ 3 | BZ 1 | BZ 2 | BZ 3 |
|--------------------|-------|-------|-------|------|------|------|
| Althererythrobacter sp. | 0     | 0     | 1     | 0    | 0    | 0    |
| Bradyrhizobium sp.   | 9     | 12    | 13    | 6    | 16   | 5    |
| Paracoccus sp.       | 0     | 1     | 0     | 0    | 0    | 0    |
| Rhizobium sp.        | 0     | 1     | 0     | 0    | 0    | 0    |
| Caballeronia sp.     | 2     | 0     | 0     | 1    | 0    | 0    |
| Duganella sp.        | 8     | 0     | 0     | 0    | 0    | 0    |
| Paraburkholderia sp. | 4     | 6     | 5     | 4    | 2    | 1    |
| Moraxella sp.        | 2     | 0     | 0     | 1    | 1    | 2    |
| Pseudomonas sp.      | 12    | 0     | 1     | 1    | 4    | 1    |
| Micrococcus sp.      | 2     | 0     | 0     | 1    | 1    | 2    |
| Nocardioides sp.     | 1     | 0     | 0     | 1    | 0    | 0    |
| Paenbacterillus sp.  | 0     | 0     | 2     | 0    | 0    | 0    |
| Unknown              | 1     | 10    | 1     | 9    | 3    | 2    |
| **Total**            | **41**| **30**| **23**| **23**| **26**| **10**|

4. Discussion

4.1. Nodulation: Does Fire Play a Role?

After fire, soils are enriched in ammonia and nitrates, as expected considering they are the inorganic forms of nitrogen produced [37], which is shown by values of N-NH$_4^+$ and N-NO$_3^-$ (mineral N forms) that almost doubled for burnt soils (see Table 1). Previous studies proposed that nodulation is downregulated through environmental feedback within the presence of ammonia, ultimately saving resources [38,39]. Additionally, Streeter [40] and Gordon et al. [41] showed that the presence of nitrates on the soil could delay nodule development. Notwithstanding, in the present study, nodulation seems to be potentiated somehow after the fire, as observed by the higher average number of nodules found in young plants growing in burnt study sites. Interestingly, through PCA biplot analysis, it can be seen that the isotopic signature of nitrogen in leaves, represented by $\delta^{15}$N_L suggests that nitrogen fixation in post-fire conditions is not correlated with the number of nodules (Figure 2b). Even considering the lower number of nodules in unburnt zones, an increased symbiotic nitrogen fixation occurred. On the other hand, there is a negative correlation between $\delta^{15}$N in leaves and mineral N forms, pointing out symbiotic nitrogen fixation dependency in the absence of mineral N forms, as it is occurring in unburnt zones (Figure 2a). With this in mind, why does A. longifolia invest in nodulation? We may hypothesize that these young plants may respond to fire events, showing a different behaviour in this “new” environment, which may be particularly relevant for plant fitness. Additionally, studies developed by Harper [42], showed that a supply of both soil and symbiotic nitrogen is required for a more favourable production of soybean. For this reason, we can possibly extend this hypothesis of partial contribution of both soil and symbiotic nitrogen to A. longifolia too, especially in the after-fire scenario that requires a faster adaptation.

Interestingly, $\delta^{15}$N in nodules is extremely enriched in both zones, indicating absence of atmospheric nitrogen fixation. One potential explanation could be ascribed to a fractionation leading to an enrichment associated with nodule development and compounds’ synthesis to export, with the second one contributing more [43]. Moreover, and in the same direction of an additional fractionation towards a higher enrichment in nodule tissue, Michelsen et al. [44] showed that $^{15}$N abundance could be influenced by the presence of mycorrhizal fungi, leading to an enriched value of the host plant up to 8%. Considering this, we can hypothesize that $\delta^{15}$N signature and the respective $^{15}$N enrichment in collected nodules could be due to an association with fungi, along with bacteria presence, which was not explored in the present study. Further studies should focus on exploring this possibility of tripartite
symbiosis already described for other *Acacia* species [45], along with nitrogen compounds exported to the plant.

The nodulation process involves the acquisition of symbiotic partners engaged in nitrogen fixation that consequently allows *A. longifolia* to access a different pool of nitrogen, facilitating recolonization. For this reason, nodulation seems to be a good process to allocate energy in, leading to its successful dominance. The higher quantity of P$_2$O$_5$ present in burnt soils (five times higher compared with unburnt soils, Table 1), is also an important factor to consider and previous studies have related it with nodules development, quantity, and function [46,47]. In the present study, we can consider it a facilitator for nodule formation in burnt zones in comparison to unburnt, due to a possible limitation in the second.

PCA biplots show a clear separation between unburnt and burnt zones, but besides this natural variability among three unburnt and three burnt zones, there is a tendency for a similar response due to ecosystem rebalance capacity after a disturbance, like a fire.

### 4.2. Nodule Bacteriome: Who Is Taking Part?

*Acacia longifolia* seems to establish symbiosis with different bacteria, beyond the commonly described rhizobia and, for the first time, our study was focused on studying bacteriome diversity in this plant species beyond nitrogen-fixers. Its “bacteriome” seems to include α-Proteobacteria, β-Proteobacteria, γ-Proteobacteria, Firmicutes and Actinobacteria. Thus, we can hypothesize that within an invasive range, *A. longifolia* can take advantage of its promiscuity, outcompeting native species, and investing energy in nodulation, considering that its ability to obtain symbiotic partners is facilitated. Furthermore, *A. longifolia* could be an example of Taylor et al.’s [48] studies who suggested that legumes, as individuals, could establish symbiosis with multiple rhizobia species simultaneously, again a direct consequence of promiscuity, making *A. longifolia* a generalist mutualist. Besides this, several studies [19,49–51] showed that *Bradyrhizobium* is the most common symbiont genus in both native and invasive range of *Acacia* species, *A. longifolia* included, which is confirmed in our study both by dominance and intragenic diversity. *Bradyrhizobium cytisi* is the main partner among this genus and it is described here for the first time as being involved in symbiosis with *A. longifolia*. On the other hand, *B. japonicum*, previously described as the major partner by Rodríguez-Echeverría [23], in both native and non-native ranges [52], was not present among our isolates. Besides other *Bradyrhizobium* species including isolates only identified to genus level, *B. canariense*, *B. ganzhouense*, *B. rifense* and *B. pachyrhizi* were also present highlighting the intrageneric diversity and genus dominance, which is in accordance with Rodriguez-Echeverría [18], who already mentions *B. ganzhouense* as present in *Acacia* nodules.

Surprisingly, we only observe three isolates belonging to *Rhizobium* genus, namely *Rhizobium rhizogenes* at species level (Supplementary Figure S1, Supplementary Table S1), which was a genus already described as one of the main symbionts among legumes [53], and particularly among *Acacia* genus in Australia (native range) [49,54]. In our study, i.e., within an invasive range, it was not so dominant. Rhizobia obtained from nodules, isolated in nitrogen-free medium as performed in our work, has been described as functional in nitrogen fixation in Leguminosae by several authors [55,56]. These authors showed that *Sinorhizobium* and/or *Mesorhizobium* related strains, isolated from *Medicago* and *Acacia*, were highly effective in nitrogen fixation (as assessed by its N$_2$ fixation effectiveness index) and induction of nodulation.

*Paraburkholderia* and *Pseudomonas* were two genera also present in our collection. *Paraburkholderia caledonica* and *Pseudomonas moorei* were already described as plant growth promoting bacteria (PGPB), playing the role of nodule inducers, presenting similarities to rhizobial species, regarding *nif* genes and *nod* factors [57]. Additionally, Saïdi et al. [58] showed that *Pseudomonas* spp. could have a role as P-solubilizer and in siderophore production. Furthermore, Martinez-Hidalgo and Hirsch [26], in their review, also highlighted the role of *Micrococcus* strains, a genus to which some isolates in our collection belong, as a plant-development “helper”. With this in mind, the question that remains is what could be the role of these non-fixing bacteria in *A. longifolia* nodulation?
In this context, recent studies postulate that the more diverse a bacterial community is within a symbiosis, the more likely it contains an effective symbiont. Such diverse symbiotic partnerships were explained by Mårtensson et al. [59] that showed that legumes cannot predict the nitrogen fixation efficiency before nodules are established and fixation is in progress; if so, we may hypothesize that *A. longifolia* emits signals that can be received by several soil bacteria. Other authors [60] have proposed that legumes can control nodulation through oxygenation of nodule microenvironment, leading to bacteria death and nodule senescence, showing that “the host controls the party” [61]. Considering this, diversity is easy to be under control. This could be *A. longifolia*’s strategy, supported by *Bradyrhizobium* spp. dominance and diversity. We can hypothesize that a process of specialization is present between *A. longifolia* and *Bradyrhizobium* spp., by comparison of unburnt and burnt zones. Thus, this symbiotic partner ensures efficient nitrogen fixation, as an obligatory partner. This great representation of *Bradyrhizobium* spp. can occur considering that different strains of the same rhizobia may differ in their effectiveness [62].

It is also known that some bacteria have functional traits that could complement each other in a way to facilitate a third functional trait [63], which is also potentiated within a wide-range community, with *Paraburkholderia* spp., *Pseudomonas* spp. and *Micrococcus* spp. presence, as possibly occurred in the present study. For this reason, further investigation should rely on the functions that could be performed by bacteria hosted in nodules, along with nitrogen fixation. In other words, nodulation would be much more than just a way to get ammonia. An interesting comparison could be carried out using Next Generation Sequencing techniques to assess a much greater diversity present inside nodules and that will complement this culture-dependent approach.

While Richardson et al. [64] suggested that mutualisms render plant species less prone to invade, our study shows that *A. longifolia* symbiosis seems to contribute for plant growth and colonization that can be due to an unspecific plant–bacteria interaction already mentioned. In fact, one of the main reasons why *A. longifolia* is such an aggressive invader, described as ecosystem-engineer, is its capacity to be eventually infected and establish relationships with a wide-range diversity of bacteria available in soils. In addition, belowground microbial diversity is substantially different after fires and this non-specific partnership can as such be useful. Besides the higher diversity in unburnt zones, some selection seems to occur in burnt zones. As reported by Franche et al. [65], diversity found within nodules’ “bacteriome” is exclusively between *nif* genes carrying bacteria and/or already described nitrogen-fixers. This specialization is also according to the nitrogen fixation pathways corroborated by the δ^{15}N in leaves ascribed to atmospheric nitrogen fixation through symbiosis (Table 2).

### 4.3. Bacteriome: What Could be Occurring?

As a costly process, nodulation has implicit a complex signal exchange, with the plant responsible for attracting bacteria possessing *nif* genes to fix nitrogen [61]. This is, as far as we know, the main goal of nodulation. This nodule “bacteriome” diversity can be explained by the inherent highly functional- and taxonomically diverse soil microbiome, along with an absence of restriction on entry into *A. longifolia* root system. Of equal importance is the process of horizontal gene transfer (HGT) that can be a determinant mechanism to facilitate this entry. In short, genes involved in signal exchange and nodulation are part of symbiotic plasmids or highly mobile “symbiotic islands”, which can be transferred easily between different bacterial species, and even genera [66]. For this reason, “bacteriome” diversity could lead more easily to effective nodulation, once beneficial bacteria can take part in nodulation, allowing *A. longifolia* to grow and spread, faster than other species, underlining the absence of competition after fires. Of course, among potential efficient nitrogen-fixers, some hitchhikers could take a ride and take advantage of nodule environment, stressing why bacteriome functionality should be explored.

Plant–bacteria interactions might, in fact, be highly regulated by environment. The observed highly efficient bacterial community inside nodules makes *A. longifolia* a top invasive species. Future studies should rely on nodule functionality, activity, and regulation, once which host “controls” the party is...
known, with guests having their own behaviour. The extent of this approach to native range could be a great contribution to understanding the invasive behaviour of *A. longifolia* and ultimately leading to its control.

5. Conclusions

In the present study, fire influenced bacterial diversity inside nodules, maintaining its nitrogen fixation functionality. After the fire disturbance, *A. longifolia* apparently “selects” nitrogen-fixing bacteria, culminating in *Bradyrhizobium* spp. dominance and intrageneric diversity. *B. cytisi* and other species in this genus seem to have a determinant role in symbiosis with *A. longifolia*, revealing a close relation and a putative facilitation role. However, besides this straight relation with *A. longifolia-Bradyrhizobium* spp., a considerable bacterial diversity was reported in our study, that could be functionally diverse and render nodules a highly complex structure.

*A. longifolia* is a typical invader that easily adapts to disturbances, and environmental changes seem to cause a different response in unburnt and burnt zones. This highlights the mutual contribution of ammonia/nitrates and symbiotic nitrogen fixation to plant development, albeit with fast ecosystem rebalance capacity.

Thus, regarding its major impacts, *A. longifolia* is not only an “ecosystem-engineer” in the aboveground environment, but also, due to its efficiency in selecting bacterial guests, it behaves as an “engineer” of the belowground environment, too.

Supplementary Materials: The following are available online at http://www.mdpi.com/1424-2818/12/6/250/s1.

Figure S1: Dendrogram based on cluster analysis of fingerprinting PCR profiles of the isolates from *A. longifolia* nodules, using the Pearson correlation coefficient and the unweighted pair-group method with arithmetic mean algorithm (UPGMA). 84% was the cut-off level below which isolates could be considered different. On the right are represented: isolate identification (CJJ xxx), zone from where it was isolated (UBZ/BZ x), Gram test result, morphology (rods (B) or cocci (CC)), catalase test result and oxidase test result, both (+) or (−). Colours are according to the phylum/class into each genus belong to: Proteobacteria/α-proteobacteria (blue), Proteobacteria/β-proteobacteria (orange), Proteobacteria/γ-proteobacteria (green), firmicutes/Bacilli (purple) and actinobacteria/Actinobacteria (yellow). Roman numbering identifies clusters. Isolates are identified up to genus level. Table S1: Identification of bacterial isolates obtained from unburnt and burnt zones by BLAST analysis of the 16S rRNA gene sequences. GenBank accession numbers are also indicated.

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