Exposure to multiple pesticides and neurobehavioral outcomes among smallholder farmers in Uganda

Samuel Fuhrimann, Andrea Farnham, Philipp Staudacher, Aggrey Atuhaire, Tiziana Manfioletti, Charles B. Niwagaba, Sarah Namirembe, Jonathan Mugweri, Mirko S. Winkler, Lutzen Portengen, Hans Kromhout, Ana M. Mora

ABSTRACT

Background: Multiple epidemiological studies have shown that exposure to single pesticide active ingredients or chemical groups is associated with adverse neurobehavioral outcomes in farmers. In agriculture, exposure to multiple pesticide active ingredients is the rule, rather than exception. Therefore, occupational studies on neurobehavioral effects of pesticides should account for potential co-exposure confounding.

Methods: We conducted a cross-sectional study of 288 Ugandan smallholder farmers between September and December 2017. We collected data on self-reported use of pesticide products during the 12 months prior to survey and estimated yearly exposure-intensity scores for 14 pesticide active ingredients using a semi-quantitative exposure algorithm. We administered 11 neurobehavioral tests to assess five neurobehavioral domains. We implemented a Bayesian Model-Averaging (BMA) approach to examine the association between exposure to multiple pesticides and neurobehavioral outcomes, while accounting for multiple testing. We applied two levels of inference to determine (1) which neurobehavioral outcomes were associated with overall pesticide exposure (marginal inclusion probability (MIP) for covariate-only models >0.5) and (2) which specific pesticide active ingredients were associated with these outcomes (MIP for models where active ingredient was included >0.5).

Results: Seventy-two percent of farmers reported use of pesticide products that contained at least one of 14 active ingredients, while the applicators used in median three different active ingredients (interquartile range (IQR) 4) in the 12 months prior to the study. The most widely used active ingredients were glyphosate (79%), cypermethrin (60%), and mancozeb (55%). We found that overall pesticide exposure was associated with impaired visual memory (Benton Visual Retention Test (BVRT)), language (semantic verbal fluency test), perceptual-motor function (Finger tapping test), and complex attention problems (Trail making A test and digit symbol test). However, when we looked at the associations for individual active ingredients, we only observed a positive association between glyphosate exposure and impaired visual memory (BVRT)([-0.24, 0] units in BVRT scores per interquartile range (IQR) increase in annual exposure to glyphosate, relative to a median [IQR] of 6 [3] units in BVRT across the entire study population). Conclusions: We found that overall pesticide exposure was associated with several neurobehavioral outcome variables. However, when we examined individual pesticide active ingredients, we observed predominantly null associations, except for a positive association between glyphosate exposure and impaired visual memory.
1. Introduction

In low- and middle-income countries (LMICs), the increase in agricultural production (OECD and FAO, 2018) has been coupled with extensive use of highly toxic pesticide active ingredients (Jepson et al., 2020; Fuhrimann et al., 2020), non-use of or inadequate use of personal protection equipment (PPE) (Fuhrimann et al., 2020; Negatu et al., 2016), and a lack of pesticide use regulations or implementation thereof (Fuhrimann et al., 2020; Jepson et al., 2014; Schreinemachers et al., 2017). Multiple studies have shown that occupational exposure to specific pesticide active ingredients or chemical groups is associated with impaired neurobehavioral outcomes (Ismail et al., 2012; Meyer-Baron et al. 2015). For example, exposure to organophosphates (OP) pesticides and carbamates – commonly assessed by measurement of acetylcholinesterase (AChE) activity – has been linked to attention problems (Rohman et al., 2016; Rothlein et al., 2006), poorer memory (Rohman et al., 2014; Roldán-Tapia et al., 2005) and motor function (Rohman et al., 2016; Roldán-Tapia et al., 2005; Starks et al., 2012) among farmers. Although few occupational studies have studied the neurobehavioral effects of pesticides active ingredients other than OP pesticides (Ismail et al., 2012; Meyer-Baron et al. 2015; Ohlander et al., 2019), a growing literature on children environmentally exposed to pesticides suggests that pyrethroid insecticides (Dalsager et al., 2019; van Wendel de Jode et al., 2016; Viel et al., 2015), manganese (Mn)-containing fungicides like mancozeb (Guner et al., 2016; Mora et al., 2018, 2016; van Wendel de Jode et al., 2016), and herbicides like glyphosate (Von Ehrenstein et al., 2019) may be associated with adverse neurobehavioral outcomes.

Exposure to multiple hazardous pesticide active ingredients is the rule rather than exception among farmers (Jepson et al., 2020). However, due to the limitations of traditional regression techniques (e.g., limited capacity to handle highly correlated exposures), previous occupational studies have relied on multiple tests of association of individual pesticide active ingredients or classes with neurobehavioral outcomes and have not accounted for co-pollutant confounding (Ismail et al., 2012; Meyer-Baron et al. 2015). Advanced variable selection and model averaging methods provide an opportunity to model the effect of correlated co-pollutant exposures (Berger et al., 2019; Hamra and Buckley, 2018). Such methods can achieve a better trade-off between false positive and negative rates than conventional analytical approaches (Agier et al., 2016; Lenters et al., 2018). Bayesian Model-Averaging (BMA) has been used in epidemiological studies to direct model selection, combined with estimation and prediction in complex multiple exposure situations. For example, BMA was recently used in a study that examined the association of phthalates and bisphenol A exposure with respiratory and allergic outcomes (Berger et al., 2019). In the present study, we applied BMA to examine the associations of complex exposure to multiple pesticide active ingredients with neurobehavioral outcomes among Ugandan smallholder farmers.

2. Methods

2.1. Study area and population

We conducted a cross-sectional study of smallholder farmers in Wakiso District, Uganda, between September and November 2017. A detailed description of the study design and methods can be found elsewhere (Dietler et al., 2020; Fuhrimann et al., 2019; Staudacher et al., 2020). Briefly, farmers were eligible to participate if they were 18 years or older and worked within the study area. To ensure a pesticide exposure contrast among study participants, we aimed to recruit an equal number (~150) of conventional farmers who used predominantly synthetic pesticides and farmers – who at least for one of their crops – used organic pest control practices. Conventional farmers were randomly sampled from lists provided by local leaders, while organic farmers were sampled from a list of organic farmers provided by a local non-governmental organization, which served as a basis for further snowballing recruiting. Over two weeks, field staff received training on tools, ethics, and research background. Additionally, we conducted a week-long pilot study with 20 farmers from the same study area. This study is nested in the Pesticide Use in Tropical Settings (PESTROP) project, which aimed to deepen the understanding of the environmental, health, and regulatory dimensions of agricultural pesticide use in Costa Rica and Uganda (Dietler et al., 2019; Winkler et al., 2019).

Table 1

Summary of pesticide application practices and exposure-intensity scores (EIS) during the 12 months prior to enrollment, Wakiso district, Uganda, 2017.

| Active ingredient | Chemical group | Type | Applicators who reported using this pesticide [n (%)] | Mixers who reported handling this pesticide [n (%)] | EIS [median (IQR)] | Yearly application [median (IQR)] | Yearly EIS-days [median (IQR)] |
|-------------------|----------------|------|-----------------------------------------------------|-----------------------------------------------------|-------------------|----------------------------------|----------------------------------|
| Total across all active ingredients |                |      | 208 (97.2)                                          | 191 (98.3)                                            | 6.1 (3.0)          | 9 (26)                           | 51.5 (179.5)                     |
| Glyphosate        | Phosphonoglycine | H    | 165 (77.1)                                          | 152 (71)                                              | 6.1 (3.1)          | 2 (2)                            | 10.0 (11.5)                      |
| Cypermethrin      | Pyrethroid      | I    | 124 (57.9)                                          | 116 (54.2)                                            | 6.4 (3.2)          | 3 (8.5)                          | 21.6 (52.8)                      |
| Mancozeb          | Bisdithiocarbamate | F   | 112 (52.3)                                          | 102 (47.7)                                            | 6.1 (3.2)          | 8 (17)                           | 46.6 (108.3)                     |
| Profenofos        | Organophosphate | I    | 105 (49.1)                                          | 99 (46.3)                                             | 6.4 (3.1)          | 2 (3)                            | 11.6 (22.1)                      |
| 2,4-D             | Alkylchlorophenoxy | H   | 97 (45.3)                                           | 92 (43)                                               | 6.1 (2.4)          | 2 (2)                            | 9.4 (11.6)                       |
| Dichlorvos        | Organophosphate | I    | 34 (15.9)                                           | 31 (14.5)                                             | 6.2 (2.7)          | 2 (3)                            | 9.2 (20.3)                       |
| Lambda-cyhalothrin | Pyrethroid     | I    | 30 (14)                                             | 27 (12.6)                                             | 5.7 (2.8)          | 2 (2)                            | 11.5 (18.6)                      |
| Dimethoate        | Organophosphate | I    | 28 (13.1)                                           | 25 (11.7)                                             | 6.4 (3.1)          | 3.5 (6)                          | 25.2 (42.5)                      |
| Chlorpyrifos      | Organophosphate | I    | 20 (9.3)                                            | 20 (9.3)                                              | 5.8 (3.8)          | 1 (0.2)                          | 6.3 (4.4)                        |
| Carabaryl         | Carbamate       | I    | 13 (6.1)                                            | 11 (5.1)                                              | 5.7 (1.6)          | 1 (1)                            | 7.1 (4.3)                        |
| Carbofuran        | Carbamate       | I    | 10 (4.7)                                            | 7 (3.3)                                               | 5.8 (4.5)          | 1.5 (2)                          | 10 (16.6)                        |
| Diazinon          | Organophosphate | I    | 10 (4.7)                                            | 10 (4.7)                                              | 5.9 (3.5)          | 1 (0)                            | 7.3 (3.6)                        |
| Paraquat          | Bipyridilium    | H    | 10 (4.7)                                            | 10 (4.7)                                              | 6.4 (1.7)          | 2 (2)                            | 14 (12.2)                        |
| Permethrin        | Pyrethroid      | I    | 10 (4.7)                                            | 10 (4.7)                                              | 6 (3.3)            | 1 (1)                            | 7.5 (3.7)                        |

1 = Insecticide, H = Herbicide, F = Fungicide.
IQR: interquartile range. *EIS between 0.89 and 13 (lowest to highest possible exposure).
Dichotomized in the regression analysis, **excluded from the regression analysis due to low number of users.
We estimated cumulative annual exposure during the 12 months prior to enrollment by combining application and mixing practices, frequency of PPE use, and hygienic behaviors after pesticide use (e.g., changing clothes and showering) via questionnaire. We also asked farmers how frequently they had used each of the 53 most commonly used pesticide products (i.e., the formulation with a specific brand name that is sold to the farmers) in the study area per month and for how many months they had used them during the 12 months preceding enrollment. These 53 pesticide products were identified based on previous surveys with smallholder farmers in Uganda (Atuhaire et al., 2017a, 2016; Clausen et al., 2017) and expertise from Uganda National Association of Community and Occupational Health (UNACOH). These products included a total of 14 pesticide active ingredients (Table 1).

2.3. Pesticide exposure assessment

We estimated cumulative annual exposure during the 12 months prior to enrollment by combining application and mixing practices, frequency of PPE use, and hygienic behavior (Eq. (1)). A detailed description of this semi-quantitative exposure algorithm can be found elsewhere (Fuhrimann et al., 2020). Briefly, exposure-intensity scores (range: 0–13) were estimated using five exposure-modifying factors. Two factors were expected to increase pesticide exposure: (i) mixing of pesticide active ingredients (MIX; score 5); and (ii) applying pesticides outdoors using manual handheld knap-sack sprayers – which was the case for all self-reported pesticide applications in our study (APPLICA- TION; score 8). Three factors were expected to decrease the exposure: (i) overall average protection achieved by PPE use, covering different body areas and accounting for differences in application frequency (PPE, score 0.14–1); (ii) time interval between pesticide application and change of clothes (CHANGE; score 0.7–1); and (iii) time interval between application and shower (SHOWER; score 0.7–1).

\[
\text{Exposure – intensity score (EIS)} = (\text{MIX} + \text{APPLICATION}) \times \text{PPE} \times \text{CHANGE} \times \text{SHOWER} \quad (1)
\]

We then estimated yearly application days per specific active ingredient by multiplying the average frequency of application per month and the number of months per year that the specific active ingredient was used. By combining the yearly application days with the EIS we estimated yearly EIS-days for all 14 pesticide active ingredients (Eq. (2)).

\[
\text{EIS – days} = \text{Pesticide active ingredient – specific yearly application days} \times \text{EIS} \quad (2)
\]

2.4. Neurobehavioral assessment

We assessed five neurobehavioral domains (i.e., language, memory, attention, executive function, and motor function) using 11 neurobehavioral tests, which resulted in 14 outcome variables (Table 2). These tests were selected because they had been used in previous studies of workers from LMICs exposed to neurotoxicants (Glass et al., 2017; Rohlman et al., 2016; van Wendel de Joode et al., 2001; Wesseling et al., 2006). Some of the tests are part of the World Health Organization (WHO) Neurobehavioral Core Test Battery (NCTB) (i.e., Benton Visual Retention Test (BVRT), Digit Span, Trail Making Test, and Digit Symbol) (Anger, 2014).

Two bilingual psychometricians (SN and JM), trained and supervised by a psychologist (TM), administered all neurobehavioral tests in the participant’s preferred language (English or Luganda) in a quiet room free from distraction.

2.4.1. Language

In the Phonetic Verbal Fluency test (Lezak, 2012), participants were asked to name as many words as possible starting with a certain initial letter (S-words in English and E-words in Luganda) in one minute. In the Semantic Verbal Fluency test (Lezak, 2012), participants were asked to notice it among four similar figures. This same procedure was repeated with nine more figures, which were progressively increasing in complexity level. We examined the number of correctly identified figures (range = 0–10) as our outcome of interest.

The Digit Span forward, a subtest of the Wechsler Adult Intelligence
Scale Revised (WAIS-R), was used to assess short-term memory (Wechsler, 1981). This test contains seven progressively longer pairs of random number sequences. These sequences were read aloud to participants and they were asked to orally repeat the digits of each sequence in the same order. Each correct sequence was scored with one point and then the total number of correct sequences (range $= 0–14$) was calculated.

2.4.3. Attention

Participants completed the Trail Making Test A, a test developed to assess attention and cognitive tracking and sequencing (Strauss et al., 2006). In this test, they were asked to join, in successive order, the numbers from 1 to 25, as fast as possible. Psychometricians pointed out errors as they occurred so that participants could immediately correct them. We examined the time (in seconds) needed to complete the test as our outcome of interest.

We also administered study participants the Digit Symbol test, a WAIS-R subtest that assesses complex attention and processing speed (Wechsler, 1981). This test required participants to match symbols to numbers according to a key located on the top of the page. The number of correctly matched symbols within 90 seconds was calculated.

The Digit Span backward (Wechsler, 1981) was used to assess working memory. In this test, seven progressively longer pairs of random number sequences were read aloud to participants and they were asked to orally repeat the digits of each sequence in reverse order. Each correct sequence in reverse order was scored with one point and then the total number of correct sequences (range $= 0–14$) was calculated.

2.4.4. Executive function

We administered the Color Trail Part 2 test (Maj et al., 1993) to assess cognitive flexibility and inhibition. Participants were asked to draw a line between numbered circles (1–25) in ascending order but alternating between two colors. We examined time (in seconds) needed to complete the test as our outcome of interest. This test was paired with the Ishihara Test (Colman, 2014) to screen for potential color blindness; however, no farmer had a deficient color vision.

The Digit Span backward (Wechsler, 1981) was used to assess working memory. In this test, seven progressively longer pairs of random number sequences were read aloud to participants and they were asked to orally repeat the digits of each sequence in reverse order. Each correct sequence in reverse order was scored with one point and then the total number of correct sequences (range $= 0–14$) was calculated.

2.4.5. Motor function

We used the Purdue Pegboard (Costa et al., 1963) to assess perceptual-motor coordination and fine motor function. Participants were asked to insert as many pegs (small rods) as possible into a row of holes on a pegboard (Lafayette Instrument Company, Model 32020A) within 30 seconds. We examined the number of pegs inserted correctly with both the dominant and the non-dominant hand as our outcome of interest.

We administered a Finger tapping test (Reed and Reed, 1997) to assess hand motoric speed. Participants were asked to press a button on an electronic tapper (Western Psychological Services, Model W-277) as many times as possible in 10 seconds for three consecutive trials. We then calculated the average number of taps for both the dominant and the non-dominant hand.

2.4.6. Exclusion criteria for neurobehavioral tests

The number of farmers excluded varied by neurobehavioral test because not all study participants completed each test and some of them had medical conditions or other circumstances that prevented them from completing the test adequately (Table 2, Fig. 1). We excluded from specific-
test analyses because they did not complete the test adequately (i.e., wrote the number sequences from the Digit span forward test on their arm instead of memorizing them) (n = 3), had eye problems (e.g., eye pain, blurry vision) (n = 4), had hand injuries or numbness (n = 6), did not understand or refused to complete a specific test (n = 15), or had incomplete entries (n = 6).

2.5. Statistical analyses

We examined differences in socio-demographic and occupational characteristics between farmers who reported applying at least one pesticide active ingredient during the 12 months preceding enrollment (henceforth called applicators, n = 208) and those who reported not applying any active ingredient in this period (henceforth called non-applicators, n = 80).

2.5.1. Analysis and selection of pesticide active ingredients

To improve our statistical power, we decided a priori to include in our analyses only pesticide active ingredients that were present in products reported to be used by at least 20 or more farmers during the 12 months prior to study enrollment. This resulted in nine active ingredients (i.e., glyphosate, cypermethrin, mancozeb, profenofos, 2,4-D, dichlorvos, lambda-cyhalothrin, dimethoate, and chlorpyrifos) out of the 14 originally assessed. We assessed the correlations between exposure estimates for these nine pesticides using Spearman correlation coefficients (\(r_s\)).

We dichotomized annual exposure to pesticide active ingredients that were used by less than 35 farmers (i.e., dichlorvos, lambda-cyhalothrin, dimethoate, and chlorpyrifos) to improve the robustness of our models and avoid independent variables with skewed statistical distributions. Non-normally distributed neurobehavioral outcomes were transformed; more specifically, time-based variables were log_{10} transformed and Digit Vigilance test scores (that included 0's) were square-root transformed (Figure S1). We used Winsorization (Kotz, 2005) to reduce the effect of outliers (defined as outside the range of the median plus or minus 3 times the interquartile range (IQR)) in the neurobehavioral outcomes’ distribution.

2.5.2. Selection of covariates

We selected our covariates a priori based on our study design (pesticide application (applicators vs. non-applicators)) and previous studies that examined occupational pesticide exposure and neuro-behavioral outcomes (Ismail et al., 2012; Meyer-Baron et al., 2015): age (19–42, 43–55, and > 55 years), education (<7th year vs. >=7th year), and psychometrician (A vs. B). We also adjusted our models for strong predictors of neurobehavioral outcomes: language of the assessment (Luganda vs. English), sex, literacy (yes vs. no), alcohol use (never vs. ever), history of head injury (never vs. ever), and HIV status (positive vs. negative).

2.5.3. Bayesian Model-Averaging (BMA) approach

We assessed associations of annual exposure to pesticide active ingredients with each neurobehavioral outcome variable using a BMA approach (R package BAS; version 1.5.5) ( Clyde et al., 2011 ). We selected the BMA approach because we expected our predictors (i.e., nine pesticide active ingredients) to spawn a very large model space that was unlikely to be dominated by any single model and was a formal response to model uncertainty (Hinne et al., 2020; Steel, 2020). Contrary to traditional all-or-none selection models, BMA computes an inclusion probability for each predictor. This inclusion probability is the sum of the posterior model probabilities over models that included this particular predictor. In BMA, the principle is that the higher the posterior model probability, the more likely it is that the model in question fits the data and therefore, in our case, that the active ingredient is in fact associated with the outcome. We used the Jeffreys-Zellner-Siow (JZS) prior for the regression coefficients and a beta-binomial (1,1) prior for the model space (Liang et al., 2008). Moreover, our BMA approach also accounted for multiple-testing across the different models (Scott and Berger, 2010).

To minimize false-positive and false-negative discovery rates, we selected exposure-outcome associations as noteworthy if the following 2-level criteria were met:

1. The sum of posterior model probabilities for covariate-only models (i.e., models with only a priori selected and/or optional covariates but no active ingredients) was <0.5 (or equivalently,
Table 4 Marginal inclusion probability (MIP) of the empty model representing the sum of posterior model probabilities for covariate-only models (i.e., models with only a priori selected and/or optional covariates but no active ingredients). Selection cut-off according to level 1 is indicated at MIP < 0.5.

| Neurobehavioral outcome | MIP |
|-------------------------|-----|
| BVRT (scores)           | 0.18|
| Finger tapping dominant hand (scores) | 0.29|
| Trail making A log10 (minutes) | 0.31|
| Finger tapping non-dominant hand (scores) | 0.42|
| Digit symbol (scores) | 0.45|
| Semantic verbal fluency (scores) | 0.50|
| Color trail making 2 log10 (minutes) | 0.68|
| Digit vigilance square root (missed 6) | 0.71|
| Digit vigilance log10 (minutes) | 0.73|
| Digit span forward (scores) | 0.74|
| Phonemic verbal fluency (scores) | 0.77|
| Digit span backwards (scores) | 0.81|
| Purdue pegboard dominant hand (scores) | 0.82|
| Purdue pegboard non-dominant hand (scores) | 0.87|

the sum of posterior probabilities in which at least one active ingredient was included was at least 0.5; and (2) the marginal inclusion probability (MIP; the sum of the posterior model probabilities for models in which the variable – or pesticide active ingredient, in our case – was included) was >0.5.

Note that (meeting of) level 2 implied (met of) level 1, but not the other way around. Level 1 was motivated by the desire to identify outcomes for which there was evidence that they were associated with exposure to at least one, but perhaps several, pesticide active ingredients, while recognizing that in some cases we might not be able to identify the specific active ingredient(s) involved. Level 2 was then used to identify the specific active ingredients involved. Therefore, we report outcomes (but no effect estimates) for which level 1 was met, whereas we report slope coefficients and 95% Bayesian Credible Intervals (BCIs) for outcomes and active ingredients that met both levels.

Statistical analyses were done in R (Foundation for Statistical Computing, version 3.6.3, RStudio version 1.2).

Table 5 Overview of main Bayesian model-averaging (BMA) estimates for six selected neurobehavioral outcome variables indicating (a) marginal inclusion probability (MIP) for the empty models (level 1) and each active ingredient (level 2), green color intensity = increasing MIP. (b) slope coefficient per interquartile range (IQR) increase in pesticide active ingredient EIS estimates for each neurobehavioral outcome variable. Orange color: pesticide active ingredient is associated with neurobehavioral outcomes for the empty models (level 1) and each active ingredient (level 2). Orange color: pesticide active ingredient EIS estimates for each neurobehavioral outcome variable. BMA models were adjusted for pesticide applicator status, age, education, and psychometrician, language of the assessment, sex, literacy, alcohol use, history of head injury, and HIV status.

| Neurobehavioral outcome | MIP | Slope coefficient per interquartile range (IQR) |
|-------------------------|-----|-----------------------------------------------|
|                        |     | BVRT (scores) | Finger tapping dominant hand (scores) | Finger tapping non-dominant hand (scores) | Digit symbol (scores) | Semantic verbal fluency (scores) | Color trail making 2 log10 (minutes) | Digit vigilance square root (missed 6) | Digit vigilance log10 (minutes) | Digit span forward (scores) | Phonemic verbal fluency (scores) | Digit span backwards (scores) | Purdue pegboard dominant hand (scores) | Purdue pegboard non-dominant hand (scores) |
| Selection level 1 (empty model) | 0.18| 0.30 | 0.31 | 0.42 | 0.45 | 0.50 | 0.64 | 0.63 | 0.63 | 0.63 | 0.63 | 0.63 | 0.63 | 0.63 |
| Selection level 2 (pesticide active ingredients) |     | (-) | (-) | (-) | (-) | (-) | (-) | (-) | (-) | (-) | (-) | (-) | (-) | (-) |
| 2,4-D | 0.15 | 0.15 | 0.08 | 0.24 | 0.09 | 0.06 | 0.014 | 0.014 | 0.014 | 0.014 | 0.014 | 0.014 | 0.014 | 0.014 |
| Chlorpyrifos | 0.08 | 0.06 | 0.46 | 0.15 | 0.26 | 0.08 | -0.008 | 0.007 | 0.013 | -0.051 | -0.272 | 0.017 | -0.001 | 0.004 |
| Cypermethrin | 0.05 | 0.06 | 0.10 | 0.05 | 0.04 | 0.10 | -0.001 | 0.004 | -0.001 | -0.002 | -0.005 | 0.022 | -0.001 | 0.004 |
| Dichlorvos | 0.07 | 0.08 | 0.22 | 0.06 | 0.12 | 0.09 | -0.008 | -0.028 | 0.005 | -0.014 | -0.100 | -0.022 | 0.005 | 0.005 |
| Dinoseb | 0.12 | 0.13 | 0.06 | 0.10 | 0.04 | 0.04 | 0.044 | 0.162 | 0.013 | 0.053 | 0.005 | 0.003 | 0.005 | 0.003 |
| Glyphosate | 0.67 | 0.48 | 0.18 | 0.24 | 0.18 | 0.08 | -0.103 | -0.217 | 0.002 | -0.087 | -0.068 | 0.007 | -0.010 | 0.031 |
| Lambda-Cyhalothrin | 0.07 | 0.08 | 0.06 | 0.06 | 0.06 | 0.10 | 0.010 | 0.031 | 0.001 | -0.022 | -0.027 | -0.031 | 0.001 | 0.001 |
| Mancozeb | 0.06 | 0.06 | 0.08 | 0.07 | 0.05 | 0.22 | 0.007 | -0.012 | -0.001 | 0.040 | -0.019 | 0.137 | -0.001 | 0.017 |
| Profenofos | 0.05 | 0.07 | 0.06 | 0.07 | 0.05 | 0.05 | -0.001 | 0.012 | 0.021 | -0.011 | -0.002 | -0.002 | -0.001 | -0.002 |

*4(+) higher scores indicate poorer performance; (-) lower scores indicate poorer performance.

3. Results

Most study participants were male (59%), were interviewed in Luganda (94%), and reported consuming alcohol (64%) (Table 3). Their median (IQR) age and years of schooling were 49 (19) years and 7 (4) years. Only a few farmers were HIV positive (6%) or reported a history of head injury (8%). Applicators were more likely to be male (70%) and literate (11%) compared to non-applicators (29% and 1%, respectively).

Seventy-two percent of farmers enrolled in our study (n = 208) were applicators and reported using pesticide products that included 14 active ingredients in the 12 months preceding enrollment (Table 1). The most widely used pesticide active ingredients were glyphosate (77%), cypermethrin (58%), and mancozeb (52%). Mancozeb was the most frequently applied active ingredient (median [IQR] application days/year 8 (17)). Most applicators (84%) reported using multiple active ingredients during the 12 months prior to enrollment (median [IQR] number of active ingredients used 3 (4)). A total of 72 different combinations of which 36 combinations were used by two or more farmers, of active ingredients were applied when considering only the nine pesticide active ingredients used by at least 20 farmers. The most frequent combination (applied by 13 farmers) consisted of cypermethrin, 2,4-D, glyphosate, mancozeb, and profenofos (Table S1).

Mancozeb had the highest EIS-days [median (IQR) 47 (108)], followed by cypermethrin [22 (53)] and dimethoate [25 (44); Table 1]. We observed strong correlations between EIS-days for glyphosate and 2,4-D (r = 0.64) and between mancozeb and cypermethrin (r = 0.63) (Fig. 2). The summary statistics for all 14 neurobehavioral outcome variables are shown in Table 2 and Figure S1.

3.1. BMA estimates

Six neurobehavioral outcome variables met our level 1 criterion (i.e., MIP of the empty model < 0.5): BVRT (summed posterior probability of covariate-only models = 0.18), Finger tapping dominant hand test (0.29), Trail making A test (0.31), Finger tapping non-dominant hand test (0.42), Digit symbol test (0.45) and Semantic verbal fluency test (0.5) (Tables 4 and 5).

According to our level 2 criterion (i.e., MIP of individual pesticide active ingredient...
Table 6
Bayesian model-averaging model (BMA) estimates for the six selected neurobehavioral outcome variables. Indicated are the marginal inclusion probability (MIP) of the empty model, the MIP for each active ingredient, its slope coefficient and the 95% Bayesian Credible Interval (BCI) per interquartile range (IQR) increase in pesticide active ingredient EIS. BMA models were adjusted for pesticide applicator, age, education, and psychometrician, language of the assessment, sex, literacy, alcohol use, history of head injury, and HIV status.

| Language | Semantic verbal fluency (scores) | Complex attention | Memory | Perceptual-motor function | Finger tapping dom. hand (scores) | Finger tapping non-dom. hand (scores) |
|----------|----------------------------------|-------------------|--------|---------------------------|----------------------------------|-------------------------------------|
|          | Empty Model/Exposure MIP Slope coef. 95% BCI | Empty Model/Exposure MIP Slope coef. 95% BCI | Empty Model/Exposure MIP Slope coef. 95% BCI | Empty Model/Exposure MIP Slope coef. 95% BCI | Empty Model/Exposure MIP Slope coef. 95% BCI | Empty Model/Exposure MIP Slope coef. 95% BCI |
| Empty Model | 0.5 NA NA NA | 0.314 NA NA NA | 0.178 NA NA NA | 0.295 NA NA NA | 0.148 NA NA NA | 0.325 NA NA NA |
| 2,4-D | 0.061 0.004 −0.004 0.036 | 0.082 0 0 0.007 | 0.151 −0.014 −0.187 3.00E− 05 | 0.295 NA NA NA | 0.148 −0.035 −0.567 3.00E− 04 | 0.235 0.438 0 2.698 |
| Chlorpyrifos | 0.078 0.017 −0.001 0.194 | 0.457 0.013 −6.00E− 06 0.043 | 0.076 −0.008 −0.121 0.007 | 0.063 0.007 −0.104 0.229 | 1.02E− 04 0.001 | 0.149 −0.051 −0.503 0.001 |
| Cypermethrin | 0.095 0.022 −0.001 0.304 | 0.096 −0.001 −0.017 0 | 0.054 −0.001 −2.00E− 04 0.007 | 0.061 0.004 0 0.149 | 0.05 −0.002 0 0.003 | 0.056 −0.014 −0.101 0.111 |
| Dichlorvos | 0.086 −0.022 −0.276 0 | 0.221 0.005 0 0.034 | 0.044 0.003 0 0 | 0.058 0 0 0 | 0.04 −0.001 0.017 | 1.00E− 04 |
| Dimethoate | 0.078 0.007 0 0 | 0.176 0.002 0 0.013 | 0.096 −0.031 −0.36 0 | 0.08 0.001 0 0.015 | 1.00E− 04 0.001 | 0.055 −0.027 −0.126 0 |
| Glyphosate | 0.078 0.007 0 0 | 0.176 0.002 0 0.013 | 0.096 −0.031 −0.36 0 | 0.08 0.001 0 0.015 | 1.00E− 04 0.001 | 0.055 −0.027 −0.126 0 |
| Lambda-Cyhalothrin | 0.096 −0.031 −0.36 0 | 0.08 0.001 0 0.015 | 0.096 −0.031 −0.36 0 | 0.08 0.001 0 0.015 | 1.00E− 04 0.001 | 0.055 −0.027 −0.126 0 |
| Mancozeb | 0.224 0.137 0 0.88 | 0.079 −0.001 −0.022 0 | 0.045 −0.002 0 0 | 0.058 0.000 0 0.003 | 0.048 −0.011 0 0 | 0.048 −0.011 0 0 |
| Profenofos | 0.045 −0.002 0 0 | 0.058 0.000 0 0.003 | 0.045 −0.002 0 0 | 0.058 0.000 0 0.003 | 0.048 −0.011 0 0 | 0.048 −0.011 0 0 |

| Empty Model | 0.061 0.004 −0.004 0.036 | 0.082 0 0 0.007 | 0.151 −0.014 −0.187 3.00E− 05 | 0.295 NA NA NA | 0.148 −0.035 −0.567 3.00E− 04 | 0.235 0.438 0 2.698 |
| Chlorpyrifos | 0.078 0.017 −0.001 0.194 | 0.457 0.013 −6.00E− 06 0.043 | 0.076 −0.008 −0.121 0.007 | 0.063 0.007 −0.104 0.229 | 1.02E− 04 0.001 | 0.149 −0.051 −0.503 0.001 |
| Cypermethrin | 0.095 0.022 −0.001 0.304 | 0.096 −0.001 −0.017 0 | 0.054 −0.001 −2.00E− 04 0.007 | 0.061 0.004 0 0.149 | 0.05 −0.002 0 0.003 | 0.056 −0.014 −0.101 0.111 |
| Dichlorvos | 0.086 −0.022 −0.276 0 | 0.221 0.005 0 0.034 | 0.044 0.003 0 0 | 0.058 0 0 0 | 0.04 −0.001 0.017 | 1.00E− 04 |
| Dimethoate | 0.078 0.007 0 0 | 0.176 0.002 0 0.013 | 0.096 −0.031 −0.36 0 | 0.08 0.001 0 0.015 | 1.00E− 04 0.001 | 0.055 −0.027 −0.126 0 |
| Glyphosate | 0.078 0.007 0 0 | 0.176 0.002 0 0.013 | 0.096 −0.031 −0.36 0 | 0.08 0.001 0 0.015 | 1.00E− 04 0.001 | 0.055 −0.027 −0.126 0 |
| Lambda-Cyhalothrin | 0.096 −0.031 −0.36 0 | 0.08 0.001 0 0.015 | 0.096 −0.031 −0.36 0 | 0.08 0.001 0 0.015 | 1.00E− 04 0.001 | 0.055 −0.027 −0.126 0 |
| Mancozeb | 0.224 0.137 0 0.88 | 0.079 −0.001 −0.022 0 | 0.045 −0.002 0 0 | 0.058 0.000 0 0.003 | 0.048 −0.011 0 0 | 0.048 −0.011 0 0 |
| Profenofos | 0.045 −0.002 0 0 | 0.058 0.000 0 0.003 | 0.045 −0.002 0 0 | 0.058 0.000 0 0.003 | 0.048 −0.011 0 0 | 0.048 −0.011 0 0 |
active ingredients >0.5), we observed that higher annual exposure to glyphosate was associated with lower Benton Visual Retention Test scores (slope coefficient per IQR increase in EIS [95% BCI] –0.10 [–0.24; 0]; MIP = 0.66) (Tables 5 and 6). We found no associations of other active ingredients with the BVRT or the other five neurobehavioral outcome variables selected according to level 1. Of note, the two highest MIPs for the selected neurobehavioral outcome variables were as follows: Finger tapping test dominant hand: glyphosate (0.48) and 2,4-D (0.15); Trail making A test: chlorpyrifos (0.46), dichlorvos (0.22); Finger tapping non-dominant hand test: glyphosate (0.24) and 2,4-D (0.15); Digit symbol test: chlorpyrifos (0.26) and glyphosate (0.17); Semantic verbal fluency test: mancozeb (0.22) and lambda-cyhalothrin (0.10). Overall, we also observed following trends for a impairment due to the following pesticide active ingredients (though not selected under level 2): dichlorvos with all six, glyphosate with five and lambda-cyhalothrin and chlorpyrifos with four neurobehavioral outcome variables.

4. Discussion

In this study of Ugandan smallholder farmers, we observed that overall pesticide exposure was associated with impaired visual memory, language, perceptual-motor function, and complex attention. However, when examining individual pesticide active ingredients, we only found an association between exposure to the herbicide glyphosate and impaired visual memory. To our knowledge, this is the first occupational study reporting an association between glyphosate exposure and a neurobehavioral outcome.

Animal studies have suggested neurotoxic effects of glyphosate and glyphosate-based herbicides, including decreased locomotor activity (Ait Bali et al., 2017; Baier et al., 2017; Gallegos et al., 2016; Martínez et al., 2018) and depressive behavior (Ait Bali et al., 2017; Cattani et al., 2017). To date, only a few human case studies have suggested that acute glyphosate exposure may lead to direct central nervous system toxicity (Malhotra et al., 2010; Potrebić et al., 2009; Wang et al., 2011). Recently, a case-control study found that prenatal residential proximity to agricultural glyphosate applications was associated with increased odds of autism spectrum disorder (ASD) during childhood (Von Ehrenstein et al., 2019). Researchers around the world have expressed their concern about the lack of data on real-world exposure to glyphosate and impaired visual memory. To our knowledge, this is the first occupational study reporting an association between glyphosate exposure and a neurobehavioral outcome.

Animal studies have suggested neurotoxic effects of glyphosate and glyphosate-based herbicides, including decreased locomotor activity (Ait Bali et al., 2017; Baier et al., 2017; Gallegos et al., 2016; Martínez et al., 2018) and depressive behavior (Ait Bali et al., 2017; Cattani et al., 2017). To date, only a few human case studies have suggested that acute glyphosate exposure may lead to direct central nervous system toxicity (Malhotra et al., 2010; Potrebić et al., 2009; Wang et al., 2011). Recently, a case-control study found that prenatal residential proximity to agricultural glyphosate applications was associated with increased odds of autism spectrum disorder (ASD) during childhood (Von Ehrenstein et al., 2019). Researchers around the world have expressed their concern about the lack of data on real-world exposure to glyphosate and impaired visual memory. To our knowledge, this is the first occupational study reporting an association between glyphosate exposure and a neurobehavioral outcome.

In this study of Ugandan smallholder farmers, we observed that overall pesticide exposure was associated with impaired visual memory, language, perceptual-motor function, and complex attention. However, when examining individual pesticide active ingredients, we only found an association between exposure to the herbicide glyphosate and impaired visual memory. To our knowledge, this is the first occupational study reporting an association between glyphosate exposure and a neurobehavioral outcome.

In this study, we found that overall pesticide exposure was associated with several neurobehavioral outcome variables in an agricultural setting in a low-income country. However, when we examined individual pesticide active ingredients, we observed predominantly null associations, except for a positive association between glyphosate exposure and impaired visual memory. Glyphosate is the most widely used herbicide in the world, but no previous studies have examined its neurobehavioral effects in farmers, while accounting for co-exposure confounding. Additional (prospective) studies are needed to replicate our findings.

5. Conclusion

In this study, we found that overall pesticide exposure was associated with several neurobehavioral outcome variables in an agricultural setting in a low-income country. However, when we examined individual pesticide active ingredients, we observed predominantly null associations, except for a positive association between glyphosate exposure and impaired visual memory. Glyphosate is the most widely used herbicide in the world, but no previous studies have examined its neurobehavioral effects in farmers, while accounting for co-exposure confounding. Additional (prospective) studies are needed to replicate our findings.

Author contributions

SF, AMM, MSW, TM, PS, CN, and AA: conception and planning. SF, PS, TM, AA, SN, and JM: collection of data. SF, AMM, LP, AF, and HK: analysis and interpretation of data. SF, LP, and AMM: drafting of the manuscript. All authors participated in editing the final version of the manuscript.

Declaration of competing interest

The authors declare that they have no conflicting interests. HK and SF report grants from CropLife Europe, outside the submitted work (IMPRESS study: www.impress-project.org).

Acknowledgments

We gratefully acknowledge the study participants and staff, namely Annah Nyesigire, Nuhuh Mutebi, Esther Mirembe, Lydia Yariwo, and Julian Nandhego. Finally, we thank the entire PESTROP team for their efforts.

In this study of Ugandan smallholder farmers, we observed that overall pesticide exposure was associated with impaired visual memory, language, perceptual-motor function, and complex attention. However, when examining individual pesticide active ingredients, we only found an association between exposure to the herbicide glyphosate and impaired visual memory. To our knowledge, this is the first occupational study reporting an association between glyphosate exposure and a neurobehavioral outcome.

In this study, we found that overall pesticide exposure was associated with several neurobehavioral outcome variables in an agricultural setting in a low-income country. However, when we examined individual pesticide active ingredients, we observed predominantly null associations, except for a positive association between glyphosate exposure and impaired visual memory. Glyphosate is the most widely used herbicide in the world, but no previous studies have examined its neurobehavioral effects in farmers, while accounting for co-exposure confounding. Additional (prospective) studies are needed to replicate our findings.
organophosphorous pesticides. Occup. Environ. Med. 60, 279–286. https://doi.org/10.1007/s11002-020-01810-z.
Fiedler, N., Kipen, H., Kelly-McNeil, K., Fenrke, R., 1997. Long-term use of organophosphates and neuropsychological performance. Am. J. Ind. Med. 32, 487-496. https://doi.org/10.1002/(SICI)1097-0290(199711)32:5<487::AID-AJIM8>3.0.CO;2-3.
Fuhrmann, S., Staudacher, P., Lindh, C.H., van Wendel de Joode, B., Mora, A.M., Winker, M.S., Kromhout, H.J., 2020. Variability and predictors of weekly pesticide exposure in applicators from organic, sustainable and conventional smallholder farms in Costa Rica. Occup. Environ. Med. 77, 40–70. doi:10.1016/0163-3438(93)90147-5.
Fuhrmann, S, Winker, M.S., Staudacher, P., Weiss, F.T., Stamm, C., Eggel, R.I.L., Lindh, C.H., Menezes-Filho, J.A., Baker, J.M., Ramirez-Munoz, F., Gutierrez-Vargaz, R., Mora, A.M., 2019. Exposure to pesticides and health effects on farm owners and workers from conventional and organic agricultural farms in Costa Rica: protocol for a cross-sectional study. JMIR Res. Protoc. 8, e10914. https://doi.org/10.2196/10914.
Gallegos, C.B., Bartos, M., Bras, C., Gumil, A., Antonelli, M.C., Minetti, A., 2016. Exposure to a glyphosate-based herbicide during pregnancy and lactation induces neurobehavioural alterations in rat offspring. Neurotoxicology 53, 20–28. https://doi.org/10.1016/j.neuro.2015.11.015.
Glaser, T., Dalvie, M.A., Holtman, A., Voert, A.A., Ramesar, R.S., London, L., 2017. DNA variants and organophosphate neurotoxicity among emerging farmers in the Western Cape of South Africa. Am. J. Ind. Med. 61, 11–20. https://doi.org/10.1002/jjem.22914.
Guiner, R.B., Arora, M., Jerrett, M., Bradman, A., Haralga, K.G., Mora, A.M., Kogut, K., Hubbard, A., Austin, C., Holland, N., Ekenazu, B., 2016. Manganese in teeth and neurodevelopment in Young Mexican-American Children. Environ. Res. 145, 688-695. https://doi.org/10.1016/j.envres.2015.09.003.
Hamm, M.J., Buckley, J.P., 2016. Environmental exposure mtures: questions and methods to address them. Curr. Epidemiol. Reports, 5, 160-165. https://doi.org/10.1007/s40471-018-0145-0.
Hinne, M., Gronau, Q.F., van der Bergh, D., Wagemakers, E.J., 2020. A Conceptual introduction to Bayesian model averaging. Adv. Methods Pract. Psychol. Sci. 3, 200–215. https://doi.org/10.1177/2525299198988657. Ismail, A.A., Bodner, T.E., Rohlman, D.S., 2012. Neurobehavioral performance among agricultural workers and pesticide applicators: a meta-analytic study. Occup. Environ. Med. 69, 457-464. https://doi.org/10.1136/oem.2011-100204.
Jepson, P.C., Guzy, M., Blautois, K., Sow, M., Minaes, P., Kegley, S., 2014. Measuring pesticide ecological and health risks in West African agriculture to establish an enabling environment for sustainable intensification. Philos. Trans. R. Soc. B Biol. Sci. 369, 20130491. https://doi.org/10.1098/rstb.2013.0491.
Jones, K., Basinas, I., Kromhout, H., et al., 2020. IMPRESS: Improving Exposure Assessment Methodologies for Epidemiological Studies on Pesticides. JMRI Res. Protoc. 9, 8-9. https://doi.org/10.1177/1664246X19851372.
Jepson, P.C., Murray, K., Bach, O., Bonilla, M.A., Neumeister, L., 2020. Selection of pesticides to reduce human and environmental health risks: a global guideline and minimum pesticides list. Lancet Planet. Health. 4, e56–e63. https://doi.org/10.1016/S2542-5196(19)30266-9.
Kotz, S., 2005. Encyclopedia of Statistical Sciences. Wiley-Interscience, New York.
Landrigan, P.J., Belpoggi, F., 2018. The need for independent research on the health effects of glyphosate-based herbicides. Environ. Heal. A Glob. Access Sci. Source. 104231. https://doi.org/10.1016/j.eh.2017.06.001.
Clausen, A., Thomsen, A., Atuhaire, A., Jors, E., 2017. Effect of integrated pest management training on Ugandan small-scale farmers. Environ. Health Insights 11. https://doi.org/10.7189/jogh.09.020201.
Lenters, V., Vermeulen, R., Portengen, L., 2018. Performance of variable selection methods for assessing the health effects of correlated exposures in case-control studies. Occup. Environ. Med. 75, 522–529. https://doi.org/10.1136/oemed-2016-104231.
Lewis, K.A., Tzilivakis, J., Warner, D.J., Green, A., 2016. An international database for pesticide risk assessments and management. Hum. Ecol. Risk Assess. 22, 1050–1064. https://doi.org/10.1080/10807039.2015.1133242.
Lewis, R., Kelland, D.Z., Kupke, T., 1990. A normative study of the repeatable cognitive perceptual-motor battery. Arch. Clin. Neuropsychol. 5, 187. https://doi.org/10.1080/088756790090095-7.
Legaz, M.D., 2012. Neuropsychological assessment. Oxford University Press, New York. Largaro, F., Molina, G., Berger, J.O., 2008. Mixtures of g priors for study of the effects of subchronic and chronic exposure to glyphosate in mice. Front. Neuro. 20, 8010. https://doi.org/10.1177/1178630217703391.
London, L., Beseler, C., Bouchard, M.F., Bellinger, D.C., Colosio, C., Grandjean, P., Vounatsou, P., Weiss, F., Wiedemann, R., Winkler, M.S., Zhou, X.-N., Utzinger, J., 2019. Health in the 2030 Agenda for Sustainable Development: from framework to action, transforming challenges into opportunities. J. Glob. Health 9, 1–6. https://doi.org/10.7189/jogh.09.020201.
Fuhrmann, S., Klinová, J., Pribilová, P., et al., 2020. Qualitative assessment of 27 current-use pesticides in air at 20 sampling sites across Africa. Chemosphere 258, 127333. https://doi.org/10.1016/j.chemosphere.2020.127333.
Farhat, T.M., Abdelraouf, G.M., Amr, M.M., Shebl, M.M., Farhat, F.M., Anger, W.K., 2003. Neurobehavioural effects among workers occupationally exposed to
