The associations of urinary DEHP metabolite levels, serum thyroid hormones, and thyroid-related genes among the adolescent students from China: a cross-sectional study

Yaming Zhao1 · Xinyue Song1 · Shuang Ding1 · Wen Qi1 · Yuezhu Zhang1 · Qi Xu1 · Tianyang Zhao1 · Xueting Zhang1 · Xu Li1 · Fuju Wu2 · Lin Ye1

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Abstract
Our study aimed to investigate the associations between DEHP exposure and serum thyroid hormone levels in 347 adolescents and young adults. We measured DEHP metabolites including mono(2-ethylhexyl) phthalate (MEHP), mono(2-ethyl-5-oxohexyl) phthalate (MEOHP), mono(2-ethyl-5-hydroxyhexyl) phthalate (MEHHP), mono(2-ethyl-5-carboxypentyl) phthalate (MECPP), and mono(2-carboxymethyl)hexyl phthalate (MCMHP) in their urine. Total thyroxine (TT4), total triiodothyronine, free triiodothyronine, free thyroxine (FT4), thyroid-stimulating hormone and the mRNA levels of thyroid peroxidase (TPO), thyroglobulin (TG), sodium iodide symporter (NIS), thyroid transcription factor 1 (TTF-1), and paired box gene 8 (PAX-8) in serum were measured. The results of statistical analysis showed that urinary DEHP metabolites were generally negatively associated with TT4 levels in serum. In the males, the FT4 levels showed positive associations with urinary MEHP, MECPP, MCMHP, and ∑DEHP. The mRNA level of TG was significantly positively correlated with the levels of MECPP, MCMHP, and ∑DEHP, while the levels of TTF-1 and PAX-8 mRNA were significantly positively correlated with the levels of DEHP metabolites. Taken together, DEHP may affect the synthesis of TG by altering the normal transcription of TTF-1 and PAX-8, leading to decreased TT4 levels in Chinese adolescents.

Keywords Adolescents · DEHP metabolites · Thyroid hormone · Human biomonitoring

Introduction
Di-2-ethylhexyl phthalate (DEHP) is the most common member of phthalates. It is used as plasticizer in many products, especially in blood bags and infusion tubes and other medical devices, and is also common in food and beverage packaging (Earls et al. 2003); architectural and furniture materials, including furniture and vinyl flooring; and perfumed cosmetics and personal care products (Dobrzynska 2016; Koniecki et al. 2011). Because of the overuse of plastic products, people can be exposed to DEHP easily. At present, a certain amount of DEHP can be detected in many environmental media (air, sediment, water, soil, etc.) (Gartner et al. 2009). The study has shown that DEHP has the endocrine, testicular, ovarian, neural, hepatotoxic, and cardiotoxic effects on animal models and humans in vitro and in vivo (Rowdhwal and Chen 2018). People are mainly exposed to DEHP through food, water, air, and intravenous fluids, among which diet is the main way (Schettler 2006). After DEHP enters the intestinal tract, it is metabolized by intestinal esterase into mono-(2-ethylhexyl) phthalate (MEHP), and then, MEHP is preferentially absorbed by the digestive tract, decomposed into mono(2-ethyl-5-oxohexyl) phthalate (MEOHP), mono(2-ethyl-5-hydroxyhexyl) phthalate (MEHHP), mono(2-ethyl-5-carboxypentyl) phthalate (MECPP), and mono(2-carboxymethyl)hexyl phthalate (MCMHP) in their urine.
The synthesis of thyroid hormone uses iodine and tyrosine as the main materials, and the synthesis process is mainly to transfer the iodine ions in the blood to the acinar cells and convert them into organic iodine under the action of NIS. Then, under the action of TPO and TG, the iodination of tyrosine residues occurs to form monoiiodine substitutions and diiodo substitutions, which are then connected to form a large amount of T4 and a small amount of T3, which are stored in the acinar space; then, under the action of thyroid stimulating hormone (TSH), it is released from the thyroid into the blood to play an active role (Di Jeso and Miccadei et al. 2002). During the anabolism of THs, the integrity of thyroid tissue, iodine absorption, and changes in related enzymes will cause abnormal thyroid function (Kloas and Lutz 2006; Shin et al. 2006).

In recent years, the incidence of thyroid-related diseases including goiter, hypothyroidism, hyperthyroidism, thyroid inflammation, and thyroid malignant tumors has continued to increase, threatening human health ([Anonymous] 2012). Endogenous factors such as age, gender, genetics, and disease can cause abnormal thyroid function (Aldossari et al. 2019; Spanou et al. 2019). Exogenous factors such as environmental factors and diet (mainly iodine intake) can also interfere with the normal metabolism of thyroid hormones (Han et al. 2019; Vigneri et al. 2017). There are increasing evidences that environmental factors such as endocrine disrupting chemical pollutants can affect thyroid hormone levels and alter thyroid morphology, which possibly lead to autoimmune thyroid diseases and thyroid tumors (Are and Shaha 2006; Huang et al. 2007; Khairy 2004; Nix et al. 2005). Epidemiological studies have found that the exposure of EEDs such as bisphenol A, PCBs, and DEHP is significantly related to the level of thyroid hormones in the body, and is closely related to the occurrence of thyroid tumors (Gorini et al. 2018; Zhao et al. 2016). As the most common environmental endocrine disruptor (EED), DEHP can interfere with thyroid hormone homeostasis and reduce thyroid hormone (TH) level, including triiodothyronine (T3), thyroxine (T4), and thyrotropin-releasing hormone (TRH) by activating the Ras/Akt/TRHR pathway, and affect hepatic enzymes which play an important role in thyroid destruction (Ye et al. 2017). Mono(2-ethylhexyl) phthalate (MEHP) is a metabolite of DEHP which also can affect the endocrine disruption of the thyroid by altering a series of genes transcription involved in the HPT axis (Zhao et al. 2014). Study has also shown that DEHP can affect TH level by controlling iodine uptake (Wenzel et al. 2005). The uptake of iodine by thyroid follicular cells depends on essential iodides such as thyroglobulin (TG), thyroid peroxidase (TPO), and crucial proteins such as sodium iodide symporter (NIS), thyroid transcription factor 1 (TTF-1), and paired box 8 (PAX-8) (Aittam et al. 2005; Targovnik et al. 2017). TTF-1 is implicated as an essential autoregulatory component in both positive and negative regulation of the TSHR and appears to be the link between TSH, the TSHR, TSHR-mediated signals, TG and TPO biosynthesis, and thyroid hormone formation (Kohn et al. 1995). In the mature thyroid, TTF-1 interacts physically with PAX-8 to induce the expression of TG and TPO (Di Palma et al. 2003; Miccadei et al. 2002). The transcription factor PAX8 is a master regulator of thyroid development and function. In the mature thyroid gland, PAX8 induces the expression of thyroid-specific genes such as those encoding TG, TPO, and NIS. The sodium (Na+)/iodide (I−) symporter (NIS) mediates iodide accumulation in the thyroid gland (Esposito et al. 1998; Fabbro et al. 1998; Ohno et al. 1999a). However, whether the expression of TG, TPO, NIS, TTF-1, and PAX-8 can be detected in blood was unclear, and it is not clear whether DEHP affects the expression of these genes in the blood.

A cross-sectional association study in the USA revealed that DEHP metabolites, total T4 (TT4), free T4 (FT4), and total triiodothyronine (TT3) were negatively correlated with thyroglobulin (TG), and were positively correlated with thyroid-stimulating hormone (TSH) (Meeker and Ferguson 2011). DEHP metabolites and mono-benzyl phthalate (MBzP) were also inversely related to total and free T3 levels, and mono-n-butylyl phthalate (MnBP) was inversely associated with total and free T3 in Danish children (Boas et al. 2010). However, most human observations were based on children and adults. Evidences of thyroid disruption among adolescents are very limited especially in Asian population. Adolescence and young adulthood are the critical periods of human growth and development, in which any change in endocrine function may cause a long-term effect on the body (Ding et al. 2019). The total intake of phthalates that did not include non-dietary intake was higher in all children than in adults (Katsikantami et al. 2016). As a result, young people are exposed to more phthalates than older people (Lin et al. 2017; Shea and Hlth 2003). Epidemiological studies have shown that phthalates may be related with premature breast development for girls, anti-androgen effects, speeding up and delay of growth and puberty in girls and boys, respectively, and poor attentional performance (Berger et al. 2018; Xie et al. 2015). In addition, some epidemiological studies have reported that phthalates may be positively associated with adolescent thyroid dysfunction (Huang et al. 2017a;
Robinson and Miller 2015; Trasande et al. 2014). This suggests that exposure to phthalates has a significant effect on the growth and development of children or adolescents and the health effects of DEHP in adolescents cannot be ignored. It is necessary to explore the association between DEHP metabolites and thyroid hormones.

Surveys of the effects of environmental endocrine disruptors’ exposure in healthy children and adolescents are valuable and may identify factors that contribute to their growth and health. Our study aimed to investigate the association of DEHP metabolites in urine with serum thyroid hormones in adolescent students. For this purpose, data was obtained from adolescent students in a university in Jilin Province. In addition, we also detected the mRNA expression of TG, TPO, NIS, TTF-1, and PAX-8 in the blood of these healthy students. The results of this study are useful to confirm the effect of DEHP on thyroid hormones and the association between them among the adolescent students.

Methods

Participants’ recruitment

This project is a cross-sectional study realized in September 2017 and in collaboration with a university in Jilin province, China, that involved newly enrolled students aged 16 to 19 in three faculties which used cluster sampling method to investigate. Excluding the patients with diseases of liver, kidney, and blood system, those receiving hormone therapy, those with family history of hereditary metabolism, and those with incomplete basic information or biological samples, a total of 347 adolescents, were finally included in the study.

This study was approved by the ethics committee of the School of Public Health, Jilin University, China. Each subject was informed of the study protocol and provided written informed consent. If they agreed to be participants of the study, a consent form would be signed by the parent/guardian or the participant aged 18 and over.

Data and sample collection

Medical history and socio-demographic information were collected using questionnaires with the parent, guardian, or participants aged 18 and over. Anthropometric measures (height and weight) and biological samples (blood and urine samples) from the participants were collected on-site at the university hospital. Before collecting samples, participants were required to be in fasting.

Urine samples were collected on-site between 8:00 and 10:00 a.m., using 15-mL polypropylene (non-plasticizer) tubes for a one-time collection of 15 mL urine samples, which were then transported frozen to the laboratory within 2 h and stored at –20 °C for further analyses. Blood samples for thyroid parameter analyses were collected in a 5-mL EDTA serum separation tube. Blood tubes were then kept at room temperature for a minimum of 30 min and a maximum of an hour, before being centrifuged at 3500 rpm for 10 min at room temperature.

Urineal DEHP metabolites determination

Urine samples were analyzed at the laboratory of Jilin Entry-Exit Inspection and Quarantine Bureau. The laboratory was assessed by the Certification and Accreditation Administration of the People’s Republic of China (CNCA) with the certificate number 150000128361.

Five DEHP metabolites in the urine of the participants were determined through solid phase extraction (SPE) combined with high performance liquid chromatography tandem mass spectrometry (HPLC–MS/MS). Analytes included secondary oxidized metabolites of DEHP, i.e., mono(2-ethylhexyl) phthalate (MEHP), mono(2-ethyl-5-oxohexyl) phthalate (MEOHP), mono(2-ethyl-5-hydroxyhexyl) phthalate (MEHHP), mono(2-ethyl-5-carboxypentyl) phthalate (MECPP), and mono(2-carboxymethyl)hexyl phthalate (MCMHP). Details of the extraction and analytical procedures are described elsewhere (Ding et al. 2019). Briefly, frozen urine samples were thawed in a 37 °C water bath and then centrifuged at 4000 rpm for 5 min. After centrifugation, the urine supernatant (1 mL) was buffered with 250 μL ammonium acetate (1 mol/L) and spiked with a 100 μL mixture of labeled internal standards of phthalate metabolites (200 ng/mL). Then, 50 μL of β-glucuronidase (250 U/mL) was added. The urine samples were incubated at 37 °C overnight in a shaking water bath. After incubation, 40 μL of formic acid and 1 mL of water were added to the urine sample to end the hydrolysis. The urine was passed by natural gravity through an HLB cartridge that had been activated by 2 mL of acetonitrile and 2 mL of 0.5% formic acid; 3 mL of 0.5% formic acid was added to the column to remove impurities. The cartridge was washed sequentially with 3 mL of acetonitrile, 3 mL of ethyl acetate, and 3 mL of chloroform. The eluate was collected in a flask and concentrated in a rotary evaporator (40 °C). The concentrated eluate was then dissolved with 1 mL of 0.1% formic acid water/acetonitrile (8:2, v/v) and sonicated for 5 min. The liquid was finally filtered through a 0.2-μm filter into the sample bottle for analysis. The detailed information including instruments, reagents, chromatographic conditions, and mass spectrometric conditions can be found in Table S1 and Table S2. The sum of DEHP metabolites (ΣDEHP metabolites) was calculated by adding the concentration of these five metabolites and expressed in nmol/L. Urinary creatinine was determined with a creatinine assay kit (Sarcosine Oxidase Method, China) using a microplate reader.
Concentrations of DEHP metabolites in urine (μg/L) were divided by creatinine concentrations in order to obtain creatinine-corrected concentrations (μg/g creatinine). Analytical results below the limit of detection (LOD) were substituted with a value of the LOD divided by the square root of 2 before statistical analysis. LODs (% > LOD) for MEHP, MEOHP, MEHHP, MECPP, and MCMHP were determined at 0.50 (100%), 0.50 (98.99%), 0.50 (98.39%), 0.50 (100%), and 0.10 (100%) μg/L, respectively.

Serum thyroid hormone determination

Measured quantification of serum levels of TT3, TT4, TSH, free T3 (FT3), FT4 using radioimmunoassay (Jiu Ding, China).

The mRNA expression of NIS, PAX-8, TTF-1, TG, and TPO determination

Total RNA was extracted from the peripheral blood of participants using Trizol Reagent (Ambion Shanghai, China). Total RNA quantity and integrity were measured with an ultraviolet spectrophotometer (IMPLEN, Germany). First-strand cDNA synthesis was carried out on total RNA by reverse transcription PCR (ABI, USA). Finally, cDNA was used for real-time quantitative PCR (Stratagene, Japan) to determine the levels of mRNAs of NIS, PAX-8, TTF-1, TG, and TPO. The primers are shown in Table S3. The mRNA expression levels were quantified via the 2−ΔCT method.

Statistical analysis

Descriptive statistics of participants’ demographics were tabulated, along with the distributions of DEHP metabolite concentrations, and thyroid hormones. The dichotomous variables are represented by N (%). The Approximate normal variables were represented by mean and standard deviation (sd), and the non-normal variables were represented by geometric mean, 95% CI, and quartiles. Thyroid hormone and DEHP metabolite concentrations were stratified according gender.

Multivariate linear regression was used to explore the associations between thyroid hormones and urinary DEHP metabolite concentrations. Concentrations of TT4 and FT3 approximated normality and were used in statistical models untransformed; the distribution of TT3, TSH and FT4 concentration was non-normality and was therefore In-transformed for statistical analyses. Creatinine-adjusted DEHP metabolite concentrations were also ln-transformed. Gender, age, BMI, and per capita income were included as covariates. BMI was modeled as a continuous variable and others as dichotomous variables. The regression coefficients were calculated as changes in thyroid hormones for an interquartile range (IQR) increase in DEHP metabolite concentrations. In addition, the associations between thyroid hormone concentrations and urinary DEHP metabolites’ levels grouped in quartiles were assessed by regression models.

Data with normal and uniform variances were analyzed by one-way analysis of variance (ANOVA), and non-normal or non-uniform data were analyzed by Kruskal-Wallis test to compare thyroid hormone-related genes mRNA expression levels. The DEHP metabolite concentrations were grouped in quartiles expressed as Q1, Q2, Q3, and Q4. Statistical significance was set at p = 0.05. Data analysis was performed and analyzed using IBM SPSS software version 24.0 (IBM, USA).

Results

Characteristics of participants

A total of 347 subjects were collected in this study, including 116 males (33.4%) and 231 females (66.6%). Age distribution ranged from 16 to 19 years old. A total of 69.7% of the subjects lived in urban and 30.3% lived in rural areas. Those whose household per capita income was less than 3,000 Yuan/month and those whose household per capita income was more than 3,000 Yuan/month accounted for 51.3% and 48.7%, respectively. The subjects’ BMI was (21.54 ± 3.67) kg/m², as shown in Table 1.

The distributions and exposure levels of DEHP in participants

The DEHP exposure level and distributions of the study subjects are shown in Table 2. In this study, the limits of the detection of urinary metabolites MEHP, MEOHP, MEHHP, MEECPP, and MCMHP were 0.50, 0.50, 0.50, 0.50, and 0.10 μg/L, respectively. Except for MEOHP (99.2%) and MEHHP (98.7%), the detection rates of all other DEHP metabolites were below the LOD.

Table 1 Demographic and socioeconomic characteristics of participants (N = 347)

| Variable          | Categories | n (%)x±s |
|-------------------|------------|----------|
| Gender            | Males      | 116 (33.4) |
|                   | Females    | 231 (66.6) |
| Age (years)       | 16~17      | 85 (24.5)  |
|                   | 18~19      | 262 (75.5) |
| Household location| Urban      | 242 (69.7) |
|                   | Rural      | 105 (30.3) |
| Income (Yuan/month)| <3000  | 178 (51.3)  |
|                   | ≥3000      | 169 (48.7)  |
| BMI (kg/m²)       | --         | 21.54 ± 3.67 |
metabolites were 100%. Among DEHP metabolites, the GM of MECPP was 9.43 μg/g CREA, and the geometric mean of other metabolites was MCMHP (5.79 μg/g CREA), MEOHP (4.55 μg/g CREA), MEHHP (3.37 μg/g CREA), and MEHP (2.43 μg/g CREA) from high to low, and the geometric mean of ∑DEHP was 94.72 nmol/g.

There were significant differences in the levels of DEHP metabolites among subjects of different ages and household location (P < 0.05). The MEOHP level in urine of 18–19 year olds was significantly higher than that of 16–17 year olds. The levels of MEOHP, MEHHP, MCMHP, and ∑DEHP in urine of subjects living in urban areas were significantly lower than those from rural areas (P < 0.05). There was no significant difference in urinary DEHP metabolites among subjects of different genders and families with per capita monthly income (P > 0.05), but the levels of DEHP metabolites in females were slightly higher than those in males except for MCMHP. Correlation analysis results among DEHP metabolites were shown in Table S1, and DEHP metabolites showed significant positive correlation (P < 0.05).

### The distributions and levels of thyroid hormone in participants

The blood thyroid hormone levels and distributions of the subjects were shown in Table 3. The serum levels of TT4 and FT3 were 124.05 ± 24.56 nmol/L and 4.57 ± 1.69 pmol/L, respectively. The geometric mean of TT3 was 1.89 nmol/L.

### Table 2 Distributions of the DEHP exposure levels among participants [GM (95% CI)]

| Variable         | MEHP (μg/g Crea) | MEOHP (μg/g Crea) | MEHHP (μg/g Crea) | MECPP (μg/g Crea) | MCMHP (μg/g Crea) | ∑DEHP (nmol/L) |
|------------------|------------------|------------------|------------------|------------------|------------------|---------------|
| Total            | 2.43 (2.23, 2.65)| 4.55 (4.15, 4.98)| 3.37 (3.06, 3.69)| 9.43 (8.70, 10.21)| 5.79 (5.30, 6.36)| 94.72 (88.50, 101.54) |
| Gender           |                  |                  |                  |                  |                  |               |
| Male             | 2.40 (2.06, 2.81)| 4.29 (3.70, 5.01)| 3.25 (2.73, 3.89)| 9.02 (7.88, 10.32)| 5.79 (5.00, 6.73)| 90.46 (79.92, 103.03) |
| Female           | 2.43 (2.20, 2.72)| 4.68 (4.17, 5.26)| 3.43 (3.03, 3.88)| 9.65 (8.71, 10.69)| 5.79 (5.17, 6.50)| 96.93 (88.70, 105.92) |
| Age (years)      |                  |                  |                  |                  |                  |               |
| 16–17            | 2.14 (1.75, 2.67)| 3.81 (3.06, 4.69)| 2.96 (2.34, 3.76)| 7.71 (6.48, 9.51)| 5.37 (4.47, 6.55)| 84.49 (72.28, 100.00) |
| 18–19            | 2.52 (2.29, 2.80)| 4.82 (4.37, 5.31) | 3.51 (3.15, 3.91)| 10.07 (9.21, 11.04)| 5.94 (5.38, 6.56)| 98.30 (90.65, 106.63) |
| Household location |                |                  |                  |                  |                  |               |
| Urban            | 2.40 (2.16, 2.68)| 4.26 (3.74, 4.76)| 3.12 (2.72, 3.53)| 8.89 (8.04, 9.87)| 5.54 (4.97, 6.21)| 89.96 (82.07, 98.34) |
| Rural            | 2.46 (2.10, 2.92)| 5.30 (4.54, 6.12) | 4.02 (3.49, 4.67)| 9.43 (8.69, 10.25)| 6.43 (5.49, 7.56)| 106.67 (93.95, 121.35)* |
| Per capita income (Yuan/month) |                  |                  |                  |                  |                  |               |
| <3000            | 2.41 (2.14, 2.72)| 4.83 (4.27, 5.50)| 3.37 (2.91, 3.88)| 9.86 (8.85, 10.98)| 6.24 (5.59, 7.02)| 98.28 (88.89, 108.34) |
| ≥3000            | 2.44 (2.13, 2.78)| 4.27 (3.73, 4.90)| 3.37 (2.88, 3.94)| 9.00 (7.88, 10.32)| 5.36 (4.71, 6.18)| 91.11 (81.18, 102.59) |

*P < 0.05.

### Table 3 Distributions of serum THs levels among participants [GM (95% CI)/ x ± s]

| Variable         | TT3 (nmol/L) | TT4 (nmol/L) | TSH (μUI/mL) | FT3 (pmol/L) | FT4 (pmol/L) |
|------------------|--------------|--------------|--------------|--------------|--------------|
| Total            | 1.89 (1.82, 1.96)| 124.05 ± 24.56| 2.56 (2.41, 2.73)| 4.57 ± 1.69| 10.01 (9.69, 10.35) |
| Gender           |              |              |              |              |              |
| Male             | 1.94 (1.83, 2.07)| 123.73 ± 24.40| 2.42 (2.19, 2.67)| 4.63 ± 1.71| 10.08 (2.43, 2.85) |
| Female           | 1.87 (1.78, 1.96)| 124.21 ± 24.69| 2.63 (2.43, 2.85)| 4.53 ± 1.68| 9.97 (2.95, 10.40) |
| Age (years)      |              |              |              |              |              |
| 16–17            | 1.84 (1.71, 1.98)| 129.07 ± 22.79| 2.48 (2.20, 2.88)| 4.40 ± 1.60| 10.12 (9.48, 10.74) |
| 18–19            | 1.91 (1.84, 2.00)| 122.43 ± 24.93* | 2.58 (2.41, 2.78)| 4.62 ± 1.72| 9.97 (9.56, 10.34) |
| Household location |            |              |              |              |              |
| Urban            | 1.91 (1.83, 2.00)| 125.07 ± 24.76| 2.60 (2.43, 2.83)| 4.48 ± 1.73| 9.99 (9.60, 10.35) |
| Rural            | 1.85 (1.76, 1.96)| 121.72 ± 24.04| 2.47 (2.24, 2.72)| 4.76 ± 1.59| 10.04 (9.32, 10.80) |
| Per capita income (Yuan/month) |             |              |              |              |              |
| <3000            | 1.89 (1.80, 1.98)| 123.41 ± 22.93| 2.42 (2.23, 2.66)| 4.56 ± 1.62| 9.98 (9.50, 10.51) |
| ≥3000            | 1.90 (1.79, 2.01)| 124.73 ± 26.21| 2.71 (2.51, 2.96)| 4.57 ± 1.77| 10.03 (9.59, 10.48) |

*P < 0.05.
that of FT4 was 10.01 pmol/L, and that of TSH was 2.56 μUI/mL.

There were significant differences in the blood TT4 levels among the subjects of different ages (P < 0.05), and there was no significant difference in serum thyroid hormone levels among different genders, family location, and family per capita monthly income (P < 0.05).

**Associations between thyroid hormones and urinary DEHP metabolites**

Since metabolites of DEHP exhibit estrogen-like activity after entering the body, subsequent analyses of this study were further stratified according to gender.

The results of multiple linear regression analysis of urinary DEHP metabolite levels and THs levels of all subjects after adjusting covariates are shown in Table 4. The levels of TT4 had a significant negative linear correlation with urine MEOHP [β = −3.599 (95% CI = −7.002, −0.625)], MEHP [β = −3.534 (95% CI = −6.774, −0.633)], MECPP [β = −4.375 (95% CI = −7.697, −0.756)], MCMHP [β = −4.397 (95% CI = −7.522, −1.174)], and ΣDEHP [β = −5.223 (95% CI = −9.880, −1.085)].

The levels of DEHP metabolites and THs of the subjects were divided into quartile groups, and the relationship between them was analyzed by ordered multi-classification logistic regression after the adjustment of covariates. The results showed that the levels of TT3 and TT4 in ΣDEHP Q4 group were at least one grade lower than those in Q1 group (P < 0.05). The levels of TT4 in Q4 groups of MEOHP, MEHP, MECPP, and MCMHP were all decreased by at least one grade compared with those in Q1 group (P < 0.05). The FT3 levels in MECPP Q3 group were significantly lower than that in Q1 (P < 0.05). The results are shown in the Fig. 1.

As shown in Table 5, there was a significant negative correlation between MEOHP and TT4 in male subjects (P < 0.05), and the levels of FT4 were significantly positively correlated with MEHP [β = 0.089 (95% CI = −0.211, 0.157)], MECPP [β = 0.106 (95% CI = 0.021, 0.190)], MCMHP [β = 0.106 (95% CI = 0.029, 0.183)], and ΣDEHP [β = 0.121 (95% CI = 0.032, 0.209)]. The results of ordered multi-classification logistic regression analysis in male subjects are shown in Fig. 2. The TT4 level of MEOHP Q4 group in male subjects was significantly lower than that of Q1 group by at least one grade. The levels of TT3 in MEHP Q3 group were significantly higher than that in Q1 (P < 0.05). Compared with Q1 group, the level of TT4 in MEHP, MECPP, and ΣDEHP Q4 group was higher. The levels of FT4 in MCMHP Q3 and Q4 groups were increased by at least one grade compared with that in Q1 (P < 0.05).

The results of multiple linear regression analysis of urinary DEHP metabolite levels and THs levels of female subjects and the ordered multi-classification logistic regression after the adjustment of covariates are shown in Table 6 and Fig. 3. In female subjects, the levels of TT4 was significantly negatively correlated with MECPP [β = −4.158 (95% CI = −8.722, 0.211)], MCMHP [β = −5.284 (95% CI = −9.133, −1.613)], and ΣDEHP [β = −5.235 (95% CI = −11.270, −0.034)]. The levels of TSH were significantly negatively correlated with ΣDEHP (P < 0.05).

The TT3 level of female subjects in ΣDEHP Q4 group was at least one grade lower than that in Q1 group (P < 0.05), and the TT3 level of female subjects in MEHP Q3 group was also significantly lower than that in Q1 group (P < 0.05). The levels of TT4 in MEOHP, MEHP, MCMHP, and ΣDEHP Q4 groups of female subjects were significantly lower than those in Q1 group (P < 0.05), and TT4 levels in MEHP and ΣDEHP Q3 groups were also significantly lower than those in Q1 group (P < 0.05). The TSH level of female subjects in MECPP and ΣDEHP Q4 group was significantly lower than that in Q1 group (P < 0.05), and the

| **Table 4** | Associations of serum thyroid hormones with urinary DEHP metabolites in total participants [β (95%CI), N=347] |
|-------------|-------------------------------------------------------------------------------------------------|
| **Total**   | **Ln TT3** | **TT4** | **Ln TSH** | **FT3** | **Ln FT4** |
| Ln MEHP    | −0.013 (−0.061, 0.043) | −1.279 (−4.620, 2.146) | −0.045 (−0.120, 0.023) | −0.042 (−0.266, 0.147) | 0.017 (−0.023, 0.058) |
| Ln MEOHP   | −0.007 (−0.066, 0.059) | −3.599 (−7.002, −0.625) | −0.039 (−0.113, 0.036) | 0.051 (−0.158, 0.256) | 0.006 (−0.032, 0.046) |
| Ln MEHP    | 0.002 (−0.045, 0.050) | −3.534 (−6.774, −0.633) | −0.035 (−0.109, 0.038) | 0.039 (−0.157, 0.227) | 0.010 (−0.026, 0.046) |
| Ln MECPP   | −0.007 (−0.058, 0.051) | −4.375 (−7.697, −0.756) | −0.020 (−0.100, 0.062) | −0.115 (−0.349, 0.116) | 0.013 (−0.031, 0.057) |
| Ln MCMHP   | −0.001 (−0.045, 0.048) | −4.397 (−7.522, −1.174) | −0.017 (−0.084, 0.051) | −0.109 (−0.314, 0.106) | 0.028 (−0.012, 0.067) |
| Ln ΣDEHP   | −0.015 (−0.076, 0.048) | −5.223 (−9.880, −1.085) | −0.054 (−0.135, 0.038) | −0.074 (−0.337, 0.185) | 0.025 (−0.025, 0.074) |

*Adjusted for age, gender, domicile, per capita income, and BMI; *P < 0.05; **P < 0.01.

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level of TSH in MEHP Q3 group was significantly lower than that in Q1 group by at least one grade (P < 0.05).

**Correlation between thyroid hormone levels and mRNA expression of key genes in thyroid hormone metabolism**

The correlation analysis results of the subjects’ blood thyroid hormone levels and mRNA expression of key genes of thyroid hormone metabolism are shown in Table 7. There was a negative correlation between FT3 level and TG mRNA expression level in blood of all subjects (P < 0.05). The TT4 level of male was significantly correlated with the mRNA relative expression levels of TG, NIS, and PAX-8 (P < 0.05). In female subjects, there was a significant negative correlation between FT3 level and TG mRNA level, and a significant negative correlation between FT4 level and NIS mRNA level (Table 7, P < 0.05).

In this study, the mRNA expression levels of thyroid hormone-related genes TPO, TG, NIS, TTF-1, and PAX-8 in blood samples were analyzed by ordered multi-classification logistic regression. In total subjects, the relative mRNA expression level of TG in MECPP and MCMHP Q4 groups was higher than that in Q1 group, and the mRNA level of TTF-1 in MECPP, MCMHP, MCMHP, and ∑DEHP Q4 groups was higher than that in Q1 group (P < 0.05). The mRNA level of PAX-8 in the MEHHP Q4 group was at least one grade higher than that in the Q1 group (Fig. 4, P < 0.05).

In male subjects, the relative mRNA level of TG in ∑DEHP Q4 group was at least one grade higher than that in Q1 group (P < 0.05). The mRNA level of TTF-1 in ∑DEHP Q2 and Q4 groups was at least one grade higher than that in Q1 group (P < 0.05). The mRNA level of PAX-8 in MEHHP
and ∑DEHP Q4 group was higher than that in Q1 group (P < 0.05). The results are shown in Fig. 5.

As shown in Fig. 6, the mRNA level of NIS of female subjects in MECPP and MCMHP Q4 groups was significantly higher than that in Q1 group (P < 0.05).

Discussion

As we all know, our study is the first research about the effects of DEHP exposure on serum thyroid hormone levels in Chinese adolescents (16–19 years). We found a negative correlation between urinary DEHP metabolites and TT4 in adolescents. When participants were stratified by gender, this association was similar in female, while FT4 levels in male were positively correlated with MEHP, MECPP, MCMHP, and ∑DEHP. Our results suggested that DEHP exposure affects thyroid hormone levels in adolescents.

Significant correlations between urinary DEHP metabolites and thyroid hormones have been reported in previous studies (Table 8). The significant correlations between the urinary DEHP metabolites (MEHHP, ∑DEHP) and TT4 observed in the Korean adult population are similar to our results (Park et al. 2017). According to the analysis results from NHANES data, metabolites of DEHP, e.g., MEHP, MEHHP, MEOHP, or MECPP, in adult urine were

![Table 6](image)

**Table 6** Associations of serum thyroid hormones with urinary DEHP metabolites in female participants [β (95%CI), N=231]

| Female | Ln TT3 | TT4 | Ln TSH | FT3 | Ln FT4 |
|--------|--------|-----|--------|-----|--------|
| Ln MEHP | −0.014 (−0.074, 0.048) | −1.232 (−5.546, 2.540) | −0.084 (−0.174, 0.008) | −0.085 (−0.356, 0.187) | −0.022 (−0.071, 0.027) |
| Ln MEOHP | 0.014 (−0.070, 0.094) | −2.583 (−7.444, 1.337) | −0.044 (−0.131, 0.038) | 0.022 (−0.230, 0.274) | −0.017 (−0.063, 0.029) |
| Ln MEHPP | 0.004 (−0.062, 0.079) | −3.345 (−7.661, 0.492) | −0.080 (−0.159, −0.005) | −0.059 (−0.295, 0.178) | −0.004 (−0.048, 0.039) |
| Ln MECPP | −0.002 (−0.071, 0.064) | −4.158 (−8.722, 0.211) | −0.086 (−0.179, 0.011) | −0.199 (−0.478, 0.080) | −0.019 (−0.070, 0.032) |
| Ln MCMHP | 0.002 (−0.052, 0.063) | −5.284 (−9.133, −1.613) | −0.045 (−0.122, 0.035) | −0.149 (−0.399, 0.100) | 0.006 (−0.040, 0.051) |
| Ln ∑DEHP | −0.011 (−0.095, 0.074) | −5.235 (−11.270, −0.034) | −0.123 (−0.231, −0.018) | −0.166 (−0.485, 0.154) | −0.014 (−0.072, 0.044) |

*Adjusted for age, domicile, per capita income, and BMI. *P<0.05. **P<0.01.
negatively correlated with TT4 (Kim et al. 2017; Meeker and Ferguson 2011). In addition, a 9-year follow-up birth cohort study in China reported that the TT4 levels were inversely associated with maternal urinary MEHHP, MEOHP, and ∑DEHP. Significant associations between urinary DEHP metabolite levels and thyroid hormones that were observed in our study are comparable to those reported previously (Huang et al. 2017b). The results were also generally consistent with experimental studies. DEHP exhibited thyroid hormone receptor antagonistic activity (Shen et al. 2009; Shi et al. 2011), and it can impair thyroid function, and significantly reduce circulatory T3 and T4 levels along with pathology changes of thyroid glands in rats (Hinton et al. 1986; Howarth et al. 2001; Liu et al. 2015; Poon et al. 1997; Price et al. 1988). In addition, DEHP can also act on thyroid follicular cells, leading to changes in thyroid hormone levels (Ghisari and Bonefeld-Jorgensen 2009; Wenzel et al. 2005). However, the patterns of the associations often varied due to the differences in participants’ characteristics, especially age and region (Boas et al. 2010; Meeker and Ferguson 2011). Our study did not find significant associations between DEHP metabolites and other thyroid hormones such as TSH and TT3. It should be noticed that the levels of DEHP metabolites in this study were lower than those in previous and reported studies (Meeker and Ferguson 2011). These differences in associations may be due to differences in DEHP exposure and thyroid hormone levels.

In this study, DEHP metabolites were detected in the urine samples of all subjects. Except for MEOHP (99.2%) and MEHHP (98.7%), all other metabolites could be detected 100%, indicating that the population was generally exposed to DEHP. We found significant differences in DEHP exposure levels among people of different ages and different household locations. The results showed that the levels of MEOHP and MECPP in the 18–19 age group were higher than those in the 16–17 age group. Although there are few age groups in this study, it can still be inferred that age may be an influencing factor affecting DEHP exposure.

Fig. 3 Regression coefficient for change in THs levels associated with quartiles change of DEHP exposure levels in female participants. Adjusted for age, BMI, household location, and per capital income, N = 231
levels. The differences in the household locations of the population also lead to differences in DEHP exposure levels. Compared with the objects whose household located in the city, those households located in the countryside have higher levels of MEOHP, MEHHP, MECPP, and $\Sigma$DEHP. This may be due to the environmental differences between the household locations. The higher level of DEHP exposure in rural areas may be related to DEHP contamination in soil and water. The rural areas are close to chemical and plastic factories. Discharges of industrial water containing DEHP can also cause soil and water pollution in rural areas (Cheng et al. 2021). In addition, plastic products such as plastic film greenhouses are widely used in agriculture and horticulture in rural China (Hu et al. 2017). DEHP in plastic films is easily released into the environment (surface water, soil, and food) and can enter the body through a variety of pathways, resulting in high levels of human DEHP exposure. Another possibility is that differences in urban and rural diets lead to urban and rural variations in DEHP exposure. Previous studies have shown that people in rural areas are more likely to eat cereals than people in urban areas (He et al. 2021), which may lead to greater exposure to DEHP in people in rural areas.

Adolescence is a recognizable stage in human development growth (Brook and Dattani 2012). Xie et al. found that MEHP exposure was significantly associated with constitutional delays in growth and puberty (CDGP) (Xie et al.

### Table 7 The correlation of THs related genes mRNA levels and THs levels

| Variable | TPO | TG | NIS | TTF-1 | PAX-8 |
|----------|-----|----|-----|-------|-------|
| Total    |     |    |     |       |       |
| TT3      | 0.086 | 0.057 | 0.068 | 0.019 | 0.077 |
| TT4      | −0.002 | 0.024 | −0.035 | −0.039 | −0.055 |
| TSH      | −0.003 | 0.016 | 0.010 | 0.023 | 0.035 |
| FT3      | −0.036 | −0.119* | 0.044 | −0.101 | −0.012 |
| FT4      | −0.030 | −0.031 | −0.093 | −0.019 | −0.036 |
| Male     |     |    |     |       |       |
| TT3      | 0.005 | −0.050 | −0.015 | −0.042 | 0.025 |
| TT4      | −0.175 | −0.185* | −0.211* | −0.129 | −0.252* |
| TSH      | 0.059 | 0.113 | 0.035 | 0.078 | 0.016 |
| FT3      | 0.057 | −0.097 | 0.074 | −0.163 | 0.014 |
| FT4      | −0.045 | −0.019 | −0.014 | 0.05 | 0.031 |
| Female   |     |    |     |       |       |
| TT3      | 0.120 | 0.107 | 0.111 | 0.045 | 0.107 |
| TT4      | 0.075 | 0.106 | 0.048 | 0.009 | 0.040 |
| TSH      | −0.076 | −0.025 | −0.001 | −0.007 | 0.041 |
| FT3      | −0.020 | −0.132* | 0.030 | −0.073 | −0.036 |
| FT4      | −0.019 | −0.019 | −0.146* | −0.036 | −0.081 |

* $P < 0.05$. 

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**Fig. 4** Regression coefficient for change in THs related genes mRNA levels associated with quartiles change of DEHP exposure levels in total participants. Adjusted for gender, age, BMI, household location, and per capital income, $N=347$
Adjusted regression coefficients in TG mRNA levels

$\Sigma$DEHP

Fig. 5 Regression coefficient for change in THs related genes mRNA levels associated with quartiles change of DEHP exposure levels in male participants. Adjusted for age, BMI, household location, and per capital income, $N = 116$

Adjusted regression coefficients in TTF-1 mRNA levels

$\Sigma$DEHP

$P_{Q2} = 0.028$

$P_{Q4} = 0.019$

Fig. 6 Regression coefficient for change in THs related genes mRNA levels associated with quartiles change of DEHP exposure levels in female participants. Adjusted for gender, age, BMI, household location, and per capital income, $N = 231$
It was also found that young girls with premature breast development had higher serum concentrations of DEHP metabolites than the control (Colon et al. 2000). Previous studies have rarely explored the association between DEHP metabolites and thyroid hormones at this stage. A US study of adolescents aged 12 to 19 found that urinary DEHP metabolites had a positive correlation with serum TT3 and TSH (Meeker and Ferguson 2011). But our research did not find this correlation. There were different results achieved in earlier life periods. Significant correlation of DEHP metabolites with FT3 were observed in Danish children aged 4 to 9 years, but not with other hormones (Boas et al. 2010). Thyroid hormone levels change significantly before adulthood (Corcoran et al. 1977). We speculated that differences in the levels of the DEHP exposure between different populations and age could be the reasons for the different effects observed. Additionally, phthalate may potentially interact with other endocrine pathways, such as the hypothalamic-pituitary axis or androgen biosynthesis. Such complex in vivo effects might be expected to contribute to differences in effects according to age. Future investigation is necessary to confirm these associations and to elucidate the biological mechanisms behind these age differences.

When stratified by gender, different associations were observed in males and females, which suggested that the association of DEHP metabolites with thyroid hormones might be gender-dependent. The Danish study on children also illustrated these gender-dependent changes, which significant associations of sum of DEHP metabolites with TT3 and FT3 were observed among girls but not boys (Boas et al.

| Country                  | Population(N) | Metabolite | T3 | T4 | FT3 | FT4 | TSH |
|--------------------------|---------------|------------|----|----|-----|-----|-----|
| China (this study)       | Students aged 16 to 20 (347) | MEOHP | ↓ |    |     |     |     |
|                          |               | MEHHP | ↓ |    |     |     |     |
|                          |               | MECPP | ↓ |    |     |     |     |
|                          |               | MCMHP | ↓ |    |     |     |     |
|                          |               | ∑DEHP  | ↓ |    |     |     |     |
| Male (116)               |               | MEHP  | ↑ |    |     |     |     |
|                          |               | MEOHP | ↓ |    |     |     |     |
|                          |               | MECPP | ↑ |    |     |     |     |
|                          |               | MCMHP | ↑ |    |     |     |     |
|                          |               | ∑DEHP  | ↑ |    |     |     |     |
| Female (231)             |               | MECPP | ↓ |    |     |     |     |
|                          |               | MCMHP | ↓ |    |     |     |     |
|                          |               | ∑DEHP  | ↓ |    |     |     |     |
| Korea (Park et al. 2017) | KoNEHS, adults aged 19 or over (6003) | MEOHP | ↓ |    |     |     |     |
|                          |               | MEHHP | ↓ |    |     |     |     |
|                          |               | ∑DEHP  | ↑ |    |     |     |     |
| USA (Kim et al. 2017)    | NHANES (1829)  | MEHHP | ↓ |    |     |     |     |
|                          |               | MEOHP | ↓ |    |     |     |     |
|                          |               | MEHHP | ↓ |    |     |     |     |
|                          |               | ∑DEHP  | ↓ |    |     |     |     |
| China (Huang et al. 2017b) | Pregnant aged 20 to 35 Boys (168) | MEOHP | ↓ |    |     |     |     |
|                          |               | MEHHP | ↓ |    |     |     |     |
|                          |               | ∑DEHP  | ↓ |    |     |     |     |
|                          |               | MEHPb | ↓ |    |     |     |     |
|                          |               |  |    |     |     |     |     |
|                          |               | MEHHP | ↑ |    |     |     |     |
|                          |               | MEOHP | ↓ |    |     |     |     |
|                          |               | MEHHP | ↓ |    |     |     |     |
|                          |               | MECPP | ↓ |    |     |     |     |
|                          |               |  |    |     |     |     |     |
|                          |               | MEHP  | ↑ |    |     |     |     |
|                          |               | MEHHP | ↑ |    |     |     |     |
|                          |               | MEOHP | ↑ |    |     |     |     |
|                          |               | MECPP | ↑ |    |     |     |     |
| USA (Meeker and Ferguson 2011) | NHANES, adults aged 20 or over (1346) | MEHP | ↓ |    |     |     |     |
|                          |               | MEOHP | ↓ |    |     |     |     |
|                          |               | MEHHP | ↓ |    |     |     |     |
|                          |               | MECPP | ↓ |    |     |     |     |
| Adolescents aged 12 to 19 (329) | | MEHP  | ↑ |    |     |     |     |
|                          |               | MEHHP | ↑ |    |     |     |     |
|                          |               | MEOHP | ↑ |    |     |     |     |
|                          |               | MEHHP | ↑ |    |     |     |     |
|                          |               | MECPP | ↑ |    |     |     |     |
| USA (Meeker et al. 2007)   | Males recruited from a study (208) | MEHP | ↓ |    |     |     |     |
| Denmark (Boas et al. 2010) | Children aged 4 to 9 (845) | ∑DEHP  | ↓ |    |     |     |     |
|                          |               | ∑DEHP  | ↓ |    |     |     |     |
|                          |               |  |    |     |     |     |     |
|                          |               |  |    |     |     |     |     |
|                          |               |  |    |     |     |     |     |
| Belgium (Dirtu et al. 2013) | Non-obese females (30) | MECPP | ↑ |    |     |     |     |
This study found that DEHP exposure in the total participants and female were negatively correlated with TT4 levels, which is consistent with the results of animal experiments (Zhang et al. 2018). Exposure to DEHP can reduce the levels of TT4 and TT3 in rats and humans. Females are more sensitive to DEHP exposure (including children and pregnant women) (Radke et al. 2019). In addition, studies have shown that thyroid hormone levels in women were more likely to change abnormally, and the probability of suffering from thyroid disease was also higher (Bauer et al. 2014). The gender-dependent associations of thyroid hormones have been reported in animal studies (Abdelouahab et al. 2008). The mechanism of gender-dependent association is not clearly explained yet, but differences in physiological and endogenous hormones between males and females may be important reasons (Abdelouahab et al. 2008; Arbuckle 2006; Hatch et al. 2010; Vaher et al. 2007). For example, estrogen can increase the concentration of TT4 in serum (Ain et al. 1987). Meanwhile, DEHP is an estrogen-like hormone (Wang et al. 2017); the androgen axis can directly affect thyroid hormones, and the antiandrogen mechanism of DEHP can destroy the balance of thyroid hormones (Flood et al. 2013). It should be noted that our study did find a positive association between DEHP metabolites and serum FT4 in males, which is contrary to previous studies (Huang et al. 2017b; Kim et al. 2017; Meeker et al. 2007). We speculate that this may be related to the differences in thyroid hormone and sex steroid hormones in female from different regions. In addition, studies have shown that male testosterone level is negatively correlated with the level of FT4 (Hassan et al. 2020), while DEHP exposure can lead to the reduction of male testosterone (Chiang et al. 2021). Therefore, DEHP exposure is positively correlated with FT4 levels in male. Another possibility is that previous studies have been reported that androgen decreases the level of serum thyroid binding globulin (TBG) (Arafah 1994; Deyssig and Weissel 1993), while DEHP has anti-androgenic effects, which may increase the levels of FT4.

The results of epidemiological and experimental studies on the effects of DEHP on the thyroid are not uniform. For population studies, gender and age may be confounding factors or influencing factors, although we adjusted them during the analysis. Therefore, when assessing the association between DEHP metabolites and thyroid hormone levels in adolescents, we collected sufficient anthropometric and adolescent-related information for the study subjects. In addition, endocrine system including thyroid hormones had a feedback system. A negative association between urinary DEHP metabolites and TT4, but no significant association with other thyroid hormones in adolescents, suggests that DEHP may alter thyroid signaling without affecting circulatory levels of thyroid hormones. Indeed, an experimental study have reported that EDs affect thyroid hormone levels independent of thyroid hormone feedback mechanism (Kaneko et al. 2008), and circulating thyroid hormone levels may not reflect the effects of EDs on thyroid hormone actions in the target tissue (Gore et al. 2015).

To better assess the association at a molecular level, we measured the mRNA expression levels of thyroid hormone-related genes TPO, TG, NIS, TTF-1, and PAX-8 in blood samples. Unlike in thyroid tissue, the expression of these genes in peripheral blood is very low. In thyroid tissue, TG has the highest expression level and is the main factor in the feedback regulation of thyroid hormone. It can regulate the growth of thyroid cells, regulate enzymes and coenzyme factors related to thyroid hormone synthesis, and participate in the synthesis and metabolism of THs. Detection of TG mRNA levels in peripheral blood leukocytes can also help distinguish benign and malignant thyroid nodules. Studies have shown that the peripheral blood TG mRNA level in the malignant thyroid nodule group was significantly higher than that in the benign nodule group. In this study, the TG mRNA level of the population was significantly positively correlated with the DEHP exposure level. We inferred the higher DEHP exposure level may be related to the occurrence of malignant thyroid nodules. The transcription factors TTF-1 and PAX-8 both play important roles in the normal development of the thyroid. PAX-8 is the main regulator of thyroid development and function. The transcription factor it encodes can participate in the growth of thyroid follicular cells and is closely related to the occurrence of thyroid cancer. In thyroid tissue, PAX-8 can induce the expression of thyroid-specific genes such as TG, TPO, and NIS (Ohno et al. 1999b). TTF-1 is a transcription factor containing a homology domain, which can regulate the expression of TPO and TG, and is considered to be an essential self-regulating component in the regulation of TSHR (Neumann et al. 2018). In this study, the mRNA levels of TTF-1 and PAX-8 were significantly positively correlated with the levels of DEHP metabolites in all subjects, which was the same as the mRNA levels of TG. Studies have shown that the synergistic effect of TTF-1 and PAX-8 can simultaneously activate the expression of TG in F133 cells (Mu et al. 2012). Therefore, we speculated that DEHP might affect the synthesis of TG by changing the normal transcription of TTF-1 and PAX-8 in peripheral blood. TG can interact with human immune cells in peripheral blood leukocytes, and may affect the levels of TT4 and TT3 in tissues and peripheral blood (Zhao et al. 2018), thus changing THs levels.

Studies have shown that TG can interact with human immune cells in peripheral blood leukocytes, and may affect the levels of T4 and T3 in tissues and peripheral blood (Zhao et al. 2018). This study also found the correlation between TG mRNA levels in peripheral blood leukocytes and thyroid hormones, indicating that DEHP exposure may affect the body’s thyroid hormone levels.
by changing the transcription of TG in peripheral blood leukocytes. NIS mediates the accumulation of iodide in the thyroid, but the role it plays in the blood is unclear. This study found that the relative levels of NIS mRNA in female subjects were different in the DEHP metabolites in the quartile groupings, and it is significantly related to the level of thyroid hormone. Although the current evidence cannot explain whether the expression of NIS in the blood has an effect on thyroid hormone, it can be inferred that NIS plays an important role in the effect of DEHP exposure on the metabolism of thyroid hormone in women. The mechanism remains to be studied.

The current research on the mechanism of DEHP exposure affecting thyroid hormone levels is mainly through animal and cell experiments. DEHP exposure can affect thyroid hormone levels by affecting the expression of key thyroid hormone enzymes and proteins, suggesting that DEHP exposure may change the normal transcription of some genes and then affect the normal secretion of thyroid hormone. However, there is no report on the relevant research on the effect of DEHP exposure on the mRNA level of key genes in thyroid hormone metabolism. The results of this study suggest that increased exposure to DEHP may affect the synthesis of TG by altering the normal transcription of TTF-1 and PAX-8, leading to decreased TT4 levels in Chinese adolescents. Certainly, this study has the same limitations as previous studies. First, this is a cross-sectional study and the number of samples is limited; we are unable to draw definitive conclusions about the association between DEHP metabolites and thyroid hormones. In addition, the level of metabolites in the urine of the participants was measured only once, which usually only reflected the short-term exposure, and is not representative of the average body burden of the subject (Cox et al. 2016), although a study has shown that a single urine sample can moderately predict an individual’s exposure level (Hauser et al. 2004; Teitelbaum et al. 2008).

**Conclusion**

Overall, the negative association between DEHP exposure and TT4 in Chinese adolescents was found and the mRNA levels of TG, TTF-1, and PAX-8 in serum changed significantly in the different quartile levels of DEHP metabolites. We speculate that increased exposure to DEHP may affect the synthesis of TG by altering the normal transcription of TTF-1 and PAX-8, leading to decreased TT4 levels in Chinese adolescents. Our results enrich the research in this field. Considering the importance of thyroid hormones, the impact of environmental factors on thyroid hormones needs further study. Additional population or mechanism studies are necessary to confirm our results.

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**Author contribution** Yaming Zhao, Lin Ye, and Juwu Fu designed the studies. Yaming Zhao, Xinyue Song, Shuang Ding, and Wen Qi performed the experiments. Yuezhu Zhang and Qi Xu analyzed the data and prepared the manuscript. Tianyang Zhao, Xueling Zhang, and Xu Li performed all the figures and tables and polished the language. All authors reviewed and approved the final manuscript.

**Data availability** The datasets used or analyzed during the current study are available from the corresponding author on reasonable request.

**Declarations**

**Ethical approval** This study was approved by the ethics committee of the School of Public Health, Jilin University, China.

**Consent to participate** Each subject was informed of the study protocol and provided written informed consent. If they agreed to be participants of the study, a consent form would be signed by the parent/guardian or the participant aged 18 and over.

**Conflict of interest** The authors declare no competing interests.
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