Short-term effects of passive restoration in springs habitats in Southern Brazil

Bruna Balestrin Piaia1,2*, Ana Paula Moreira Rovedder1,3, Idaiane Fátima Giacomini1,2, Roseline Marostega Felker1, Maureen de Moraes Stefanello1,5, Betina Camargo1,5, Djoney Procknow1,5, Jéssica Puhl Croda1,2, and

1Núcleo de Estudos e Pesquisas em Recuperação de Áreas Degradadas (NEPRADE), Universidade Federal de Santa Maria (UFSM), Santa Maria, RS, Brasil. *Corresponding author. E-mail: brunahpiaia@gmail.com.
2Programa de Pós-graduação em Engenharia Florestal, Universidade Federal de Santa Maria (UFSM), Santa Maria, RS, Brasil.
3Departamento de Ciências Florestais, Universidade Federal de Santa Maria (UFSM), Santa Maria, RS, Brasil.
4 Instituto de Meio Ambiente de Santa Catarina (IMA), Florianópolis, SC, Brasil.
5Programa de Pós-graduação em Engenharia Agrícola, Universidade Federal de Santa Maria (UFSM), Santa Maria, RS, Brasil.

ABSTRACT: This study analyzed natural regeneration floristic patterns in spring habitats submitted to different ecological restoration actions in the seasonal forest in Atlantic Forest biome, Rio Grande do Sul. We assessed floristic composition in six springs before restoration actions and 12 months later: We identified and counted all regenerating individuals with height greater than or equal to 30 cm and diameter at breast height less than or equal to 5 cm in eight plots in each spring. The richness and abundance of each spring and assessment were compared by the Kruskal-Wallis test (P<0.05). Detrended Correspondence Analysis (DCA) was performed to verify floristic patterns and association between species. Results showed an increase in richness and abundance after ecological restoration implementation. The DCA showed a distinct floristic composition between springs. The springs remaining natural vegetation and the successional stage of these vegetation influenced the natural regeneration floristic composition pattern. Passive restoration, which was effective by enclosure, favored natural regeneration recruitment.

Key words: permanent preservation area, ecological restoration, floristic composition, passive restoration.

RESUMO: O objetivo deste estudo foi analisar os padrões florísticos da regeneração natural em habitats de nascentes submetidos a diferentes ações de restauração ecológica na floresta estacional do bioma Mata Atlântica, Rio Grande do Sul. A composição florística foi avaliada em seis nascentes antes das ações de restauração e 12 meses depois. Todos os indivíduos regenerantes com altura maior ou igual a 30 cm e diâmetro à altura do peito menor ou igual a 5 cm foram contados e identificados em oito parcelas em cada nascente. A riqueza e abundância de cada nascente em cada avaliação foram comparadas pelo teste de Kruskal-Wallis (P<0,05). A Análise de Correspondência Retificada (DCA) foi realizada para verificar os padrões florísticos e a associação entre as espécies. Os resultados mostraram um aumento na riqueza e abundância após a implementação da restauração ecológica. A DCA apresentou uma composição florística distinta entre as nascentes. A vegetação natural remanescente nas áreas de nascente e o estágio sucessional da vegetação influenciaram o padrão de composição florística. A restauração passiva, efetivada pelo cercamento, favoreceu o desenvolvimento da regeneração natural.

Palavras-chave: área de preservação permanente, restauração ecológica, grupos florísticos, restauração passiva.

INTRODUCTION

A spring is where groundwater discharges to the ground surface creating a visible flow. Springs are among the most valuable and ecologically interesting ecosystems. Protection of springs is not only a question of protecting the physical feature defined as the spring, but it is also essential to preserve its surroundings to provide ecosystem functions and services, manly water quality. These fragile ecosystems are of great importance to regional biodiversity, but they are currently threatened by groundwater abstraction and pollution, habitat degradation, land use and global warming.

The degradation of forest formations in spring habitats leads to a loss of habitats and...
connectivity (BOESING et al., 2018), and a loss of biodiversity and ecosystem services (ROVEDDNER et al., 2016). As these ecosystem services, we highlighted the control of surface erosion, river flow regulation, biogeochemical cycling, microclimate and improvement of water quality (MEA, 2005). So, the conservation and ecological restoration of these habitats are very important. In addition, the Brazilian legislation defines a minimum radius of 50 meters to preserve these habitats called Permanent Preservation Area (PPA) (BRAZIL, 2012). When the spring was degraded the legislation requires the restoration to return to ecological function and ecosystem service.

In this sense, ecological restoration provides solutions to these environmental problems because it enables structure and functionality reconstruction of biological communities (MARTIN, 2017). Ecological restoration methods and actions must be adopted in order for ecological processes to return. These methods and actions must end the impact sources and provide conditions for establishing propagules so that the natural cover returns and the site has the capacity to restore its functions (NERY et al., 2013).

Among the ecological restoration methods, the passive restoration is one of lower cost approaches which uses the local regeneration potential and presents better results in landscapes with a high level of habitat connectivity (CROUZEILLES et al., 2017). At the other extreme, there are intensive strategies such as planting seedlings in a total area, called active restoration (RODRIGUES et al., 2009; RODRIGUES et al., 2011). Nucleation strategies such as planting in dense nuclei are an intermediate alternative with less intervention (BECHARA et al. 2016).

There is a general lack of scientific information regarding the specificities of the restoration methods mentioned above in southern Brazil, which has a subtropical climate (ROVEDDER et al., 2014), precisely in spring habitats. The suppression history and conversion of natural sites, as well as studies on the potential for natural regeneration (MMA, 2017) in Atlantic forest have showed the urgency of policies and programs based on ecological restoration principles.

In this context, the objective of this study was to analyze natural regeneration floristic composition patterns of springs habitats under ecological restoration by different restoration actions in the Atlantic Forest, Rio Grande do Sul.

MATERIALS AND METHODS

The study was carried out in six springs in the sub-hydrographic basin of Arroio Manoel Alves in the central region of Rio Grande do Sul (Figure 1). This sub-hydrographic basin is inserted in the Atlantic Forest and is part of the Ecological Corridor of the Fourth Colony, an important remaining of the Seasonal Forest (IBGE, 2012). There is also Mixed Ombrophilous Forest in small proportions (MARCHIORI, 2002). The relief is undulating with Cambisol and Litholic Neossol predominating in association with Regolithic Neossol soils (STRECK et al., 2008). The region’s climate is Cfa humid subtropical according to the Köppen climate classification, with hot summers and without a defined dry season. The average temperature of the hottest month is 22 °C, the coldest month is 12.2 °C and the average altitude is 400 m (ALVARES et al., 2013).

The sub-hydrographic basin of Arroio Manoel Alves had its riparian forests strongly impacted by anthropic actions regarding its use and soil occupation and the suppression of native vegetation around many springs. The main disturbance sources in the Permanent Preservation Areas (PPAs) of the springs were the suppression of native vegetation cover and replacement by Eucalyptus sp. forestry and cattle raising (Table 1), with Eucalyptus sp. forestry being the main economic activity in the sub-hydrographic basin. The six springs evaluated were fenced and received different restoration approaches in 2014, with the exception of the N6 spring, which was fenced 8 years earlier (Table 1). All six springs were perennial and punctual. The fenced area of each spring site was 0.78 ha and the characteristic of six springs assessed were:

- N1 – applied nucleation: Spring habitat with approximately 10 m radium of native vegetation in medium successional stage around the spring.
- N2 – total area seedling planting: Spring habitat with Eucalyptus sp. forestry in the rest of the area with cattle free access. This site was isolated with a fence, the Eucalyptus sp. trees were removed and seedlings were planted in nuclei. We implanted 40 nuclei of 5 seedlings each, randomly containing the following species: Eugenia involucrata DC., Schinus terebentifolia Radd., Allophylus edulis (A.St.-Hil., Cambess. & A. Juss.) Radlk., Luehea divaricata Mart. & Zucc., Campomanesia xanthocarpa O.Berg, Cupania vernalis Cambess., Prunus myrtifolia L.(Urb.), Eugenia uniflora L., Parapiptadenia rigida (Benth.) Brenan and Psidium cattleianum Sabine.
- N3 – total area seedling planting: Spring habitat with Eucalyptus sp. forestry and pasture

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that cattle had free access. This site was isolated with a fence and had seedling planting in total area with 2 x 3 distance between seedlings of the following species: *Luehea divaricata* Mart. & Zucc., *Parapiptadenia rigida* (Benth.) Brenan, *Psidium cattleianum* Sabine, *Schinus terebentifolius* Raddi, *Cupania vernalis* Cambess., *Eugenia involucrata* DC., *Allophylus edulis* (A.St.-Hil., Cambess. & A. Juss.) Radlk., *Campomanesia xanthocarpa* O.Berg, *Blepharocalyx salicifolius* (Kunth) O. Berg, *Prunus myrtifolia* L. (Urb.), *Eugenia uniflora* L. and *Vitex megapotamica* (Spreng.) Moldenke.

N3 – passive restoration: Degraded spring habitat with *Eucalyptus* sp. forestry and pasture with cattle free access. This site was isolated with a fence.

N4 – passive restoration: Disturbed spring habitat with native forest in medium successional stage with free cattle access. The site was fenced to remove cattle access.

N5 – passive restoration: Disturbed spring habitat with native forest in medium to advanced successional stage in half part of PPA zone and *Eucalyptus* sp. forest in the other half part. Cattle had free access so the site was fenced.

N6 – eight years passive restoration: Spring habitat isolated by fence eight years ago in earlier to medium successional stage. The site had *Pinus* sp. forest and cattle free access before isolation.

The vegetation data was assessed in eight plots of 2 m x 2 m in each spring systematically installed in the Permanent Preservation Area (PPA) and forming a cross-section (FLOSS et al. 2018). Four plots were installed at five meters from the spring in each of the four cardinal directions (north, south, east, west) and another four plots at 30 meters from the spring, arranged at an angle of 45° in relation to other plots. All regenerating shrub and arboreal individuals with a height greater than or equal to 30 cm and a diameter at breast height (DBH) less than or equal to 5 cm were counted and identified.

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We carried out the first assessment before implementing the restoration actions (September 2014) and the second assessment was carried out 12 months later (October 2015). The species were identified in the field and with specialists’ help from the Herbarium of the Department of Forest Sciences, Federal University of Santa Maria. The identification of botanical families followed *Angiosperm Phylogeny Group* IV (APG IV, 2016).

Shannon’s Diversity (H’) and Pielou’s Evenness (J’) indexes (MAGURRAN, 2013) were calculated using PAST software, version 1.79, and then the Shannon’s Diversity (H’) results for each spring and each assessment were compared using the Hutcheson t-test (HUTCHESON, 1970). The richness and abundance from each spring and each
assessment were compared using the Kruskal-Wallis test (P<0.05). Significant differences between springs and assessment were compared by the t-test (P<0.05) using the R software package ‘agricolae’ (MENDIBURU, 2019; R Development Core Team, 2019). The floristic composition similarity of the springs was verified by Detrended Correspondence Analysis (DCA), carried out using the CANOCO 4.5 statistical software. DCA used a species matrix for each assessment containing data on the presence and absence in each plot.

RESULTS AND DISCUSSION

A total of 1039 individuals from 29 botanical families and 76 species were verified in six springs in the two assessments. Five species were identified only in genera and eight species could not be identified. We verified 420 individuals belonging to 20 families and 46 species in the six springs studied before isolating and implementing the restoration approaches in 2014. Then after 12 months of the restoration actions, 620 individuals, 30 families and 67 species were sampled in 2015. Myrtaceae was the family with the largest number of species (14), followed by Fabaceae (5), Sapindaceae (4) and Euphorbiaceae (4). The Myrtaceae family has a large number of arboreal representatives in all phytogeographic types in Rio Grande do Sul. This family is known for its high representation in the arboreal communities of the seasonal forest, as well as Fabaceae, Euphorbiaceae and Sapindaceae (FÁVERO et al., 2015, SCIPIONI et al., 2009). Myrtaceae has an important role in the diet of frugivorous birds (CARNEVALE & MONTAGNINI, 2002). They attract seed dispersing fauna, which is important for ecological restoration.

The species with the highest occurrence were: *Blepharocalyx salicifolius* (Kunth) O.Berg, *Eugenia uniflora* L., *Calyptranthes concinna* DC, *Daphnopsis racemose* Griseb., *Miconia hiemalis* A.St.-Hil. & Naudin ex Naudin, *Prunus myrtifolia* (L.) Urb., *Allophylus guaraniticus* (A. St.-Hil.) Radlk., *Myreceugenia myricoides* (Cambess.) O. Berg, *Gymnanthes klotzschiana* Müll.Arg. and *Sebastiania brasiliensis* Spreng. (Supplementary document). The most abundant species was *B. salicifolius* reported in all springs, except for N3 in both assessments, followed by *E. uniflora* and *D. racemosa*.

There was a significant increase in richness and abundance in N4 (Table 2), this spring had native forest in medium successional stage, so passive restoration with removing cattle free access allowed the increase. In addition, we verified a significant increase in abundance in N3 (Table 2), which was the spring in more degraded conservation state. In other words, the decision for passive restoration effected by...
Table 2 - Floristic diversity indexes for the six springs under restoration, in 2014 and 2015 assessments, Arroio Manoel Alves sub-hydrographic basin, southern Brazil.

|               | 2014 | 2015 |
|---------------|------|------|
|               | N1   | N2   | N3   | N4   | N5   | N6   | N1   | N2   | N3   | N4   | N5   | N6   |
| Richness’     | 19cd | 18bc | 3e   | 21bc | 17bc | 21bc | 22bc | 21bc | 11d  | 29a  | 33bc | 29ab |
| Families      | 11   | 10   | 2    | 12   | 10   | 14   | 14   | 12   | 10   | 16   | 16   | 16   |
| Abundance*    | 71cd | 82abc| 4e   | 66cd | 10abc| 86bc | 103bc| 10abc| 33d  | 162a | 91bc | 128ab|
| H’**          | 2.41ab** | 2.47ab | 1.04 | c    | 2.76a | 2.40b | 2.76a | 2.70ab | 1.73c | 2.85a | 3.07a | 2.96a |
| J’            | 0.62 | 0.62 | 0.94 | 0.75 | 0.65 | 0.75 | 0.75 | 0.68 | 0.67 | 0.51 | 0.59 | 0.66 |

Where: N1: spring habitat in applied nucleation; N2: total area seedling planting; N3: degraded spring habitat in passive restoration; N4: disturbed native vegetation in passive restoration; N5: disturbed spring habitat in passive restoration; N6: spring habitat in passive restoration for eight years; H’: Shannon index; J’: Pielou’s Evenness Index. *Richness and Abundance followed by the same letter for springs and assessment do not differ by the Kruskal-Wallis test (P<0.05). ** Shannon’s index followed by the same letter for springs and assessment does not differ from each other by the Hutcheson t-test (P<0.05).

...the fencing and exclusion of the degradation factor (cattle) could contribute to the expression of natural regeneration in the short-term after the restoration actions. Passive restoration can favor biodiversity recuperation in areas of environmental liability, as long as there are adequate resilience levels helping to increase connectivity in fragmented landscapes (ROVEDDER et al., 2018).

The N4 (disturbed native vegetation in passive restoration), N5 (disturbed spring habitat in passive restoration) and N6 (eight years passive restoration) springs showed the highest floristic diversity in both assessments (Table 2). Moreover, the N5 spring showed a significant increase in diversity in 12 months (Table 2). These results can be attributed to local resilience with the greater range of native vegetation surrounding these springs (Table 1), which works as a propagule source and facilitates the circulation of seed dispersers. The longer time under passive restoration (eight years) in N6 may have contributed to greater diversity, while N3 presented the lowest floristic diversity in both assessments, confirming the degradation state of this spring due to pasture and Eucalyptus sp. forest.

The Detrended Correspondence Analysis (DCA) for the assessment carried out in 2015 showed a new floristic composition pattern (Figure 3). The eigenvalue for the first axis was 56.93%, summarizing more than half of the data variation. The N6 approached N1 and N4 springs in 2015. N5 showed a distinct floristic composition and did not approach any other spring in 2015, as we verify in 2014 assessment. The fact that N5 (disturbed spring habitat in passive restoration) showed a distinct floristic composition in both assessments may be related to a native forest in medium to...
advanced successional stage in half part of PPA zone, which contributed to resilience levels. We can highlight some of climax shade tolerant and climax light tolerant species that occur exclusively in N5 in 2015: *Trichilia elegans* A. Juss., *Campomanesia xanthocarpa* O. Berg, *Cedrela fissilis* Vell., *Banara tomentosa* Clos, *Eugenia hiemalis* Cambess., *Eugenia uruguayensis* Cambess and *Helietta apiculata* Benth.

The spring under passive restoration for eight years (N6) showed association with numerous species, some of which are exclusive to this spring, which possibly influenced that the first assessment did not show similarity with any other springs. We can highlight *Escallonia bifida* Link & Otto, *Cupania vernalis* Cambess. and *Mimosa* sp. as exclusive species. *C. vernalis* is one of the main species to compose the seed bank and the regenerative stratum under shade in seasonal forests (GIEHL et al., 2007), which demonstrated the effectiveness of the passive restoration effected by the fence for the return of conditions to species establishment in N6. DCA showed the species *Schinus terebinthifolia* Raddi, *Matayba elaeagnoides* Radlk., *Prunus myrtifolia* (L.) Urb., *Styrax leprosus* Hook. & Arn., *Cestrum euanthes* Schltdl. and *Bernardia pulchella* (Baill.) Müll. Arg., more associated with N6 spring in 2015 assessment.

In relation of invasion exotic species, *Hovenia dulcis* Thunb. and *Ligustrum lucidum* W. T. Aiton occurred exclusively in N6 (eight years passive restoration) and *Citrus* sp. occurred exclusively in N2 (total area seedling planting). Although, the density of individuals for these species was low; this fact deserves attention since biological invasion by exotic species is one of the main problems of passive restoration (PRACH et al., 2019). The interpretation of resilience and biological invasion levels must be realized carefully. Studies in the same region have highlighted the issue of biological invasion, with emphasis on the presence of *H. dulcis*, *L. Lucidum* (HUMMEL et al. 2014; ROVEDDER et al., 2018).

DCA showed association with *Celtis iguanaeae* (Jacq.) Sarg., *Lonchocarpus campestris* Mart. ex Benth. and *Erythroxylum deciduum* A.St-Hil. for N2 spring in 2015 assessment and for 2014 assessment we verify *Ilex brevicuspis* Reissek.
and *Rubus brasiliensis* Mart. Despite the applied nucleation in N1 and total area seedling planting in N2, the main effect observed in the short term was due to the enclosure and cattle removal in these springs. N3 spring presented association with *Solanum mauritianum* Scop, *Miconia hyemalis* A.St.-Hil. & Naudin ex Naudin, and *Machaerium paraguariense* Hassl in 2015, while in 2014 we did not verify association with species in this spring. The *Miconia* and *Solanum* genera are pioneer shrubs with a short life cycle and they colonize large clearings (TABARELLI & MANTOVANI, 1999), confirming the initial successional stage of this spring.

The composition of forest species in early communities is a reflection of local colonization patterns and the effects of remaining vegetation (CHAZDON et al., 2007). The positive results in increasing diversity and density, even in a short term of 12 months, reflect the presence of the remaining vegetation around some springs and corroborate the Ministry of the Environment’s classification for the potential of natural regeneration of the Atlantic Forest in same region (SANSEVERO et al., 2018).

CONCLUSION

The floristic composition of springs habitats was singular and distinct from each disturbed and degraded source. The local resilience and the remaining vegetation in different states of successional stage in the springs influenced the natural regeneration floristic composition patterns. Passive restoration carried out by fencing and cattle removing favored natural regeneration tree recruitment in the short term.

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DECLARATION OF CONFLICT OF INTERESTS

The authors declare no conflict of interest. The founding sponsors had no role in the design of the study; in the collection, analyses, or interpretation of data; in the writing of the manuscript, and in the decision to publish the results.

AUTHORS’ CONTRIBUTION

All authors contributed equally to the manuscript.

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