Predicting the potential for natural recovery of Atlantic salmon (Salmo salar L.) populations following the introduction of Gyrodactylus salaris Malmberg, 1957 (Monogenea)

Denholm, SJ; Hoyle, AS; Shinn, AP; Paladini, G; Taylor, NGH; Norman, RA

Published in:
PLoS ONE

DOI:
10.1371/journal.pone.0169168

First published: 29/12/2016

Document Version
Publisher's PDF, also known as Version of record

Link to publication

Citation for published version (APA):
Denholm, SJ., Hoyle, AS., Shinn, AP., Paladini, G., Taylor, NGH., & Norman, RA. (2016). Predicting the potential for natural recovery of Atlantic salmon (Salmo salar L.) populations following the introduction of Gyrodactylus salaris Malmberg, 1957 (Monogenea). PLoS ONE, 11(12), [e0169168].
https://doi.org/10.1371/journal.pone.0169168

General rights
Copyright and moral rights for the publications made accessible in the public portal are retained by the authors and/or other copyright owners and it is a condition of accessing publications that users recognise and abide by the legal requirements associated with these rights.

• Users may download and print one copy of any publication from the public portal for the purpose of private study or research.
• You may not further distribute the material or use it for any profit-making activity or commercial gain
• You may freely distribute the URL identifying the publication in the public portal

Take down policy
If you believe that this document breaches copyright please contact us providing details, and we will remove access to the work immediately and investigate your claim.

Download date: 11. May. 2021
Predicting the potential for natural recovery of Atlantic salmon (*Salmo salar* L.) populations following the introduction of *Gyrodactylus salaris* Malmberg, 1957 (Monogonea)

Scott J Denholm¹,²*, Andrew S Hoyle²†, Andrew P Shinn³, Giuseppe Paladini³, Nick G H Taylor⁴ and Rachel A Norman²,³

¹Integrative Animal Sciences, Animal & Veterinary Sciences Research Group, Scotland’s Rural College (SRUC), Edinburgh, UK.

²Computing Science and Mathematics, School of Natural Sciences, University of Stirling, Stirling, UK.

³Institute of Aquaculture, School of Natural Sciences, University of Stirling, Stirling, UK.

⁴Centre for Environment, Fisheries & Aquaculture (CEFAS), Weymouth Laboratory, Weymouth, UK.

*Corresponding author

Email: scott.denholm@sruc.ac.uk (SJD);

†These authors also contributed equally to this work
Abstract

*Gyrodactylus salaris* (Monogenea, Platyhelminthes) is a notifiable freshwater pathogen responsible for causing catastrophic damage to wild Atlantic salmon stocks, most notably in Norway. In some strains of Baltic salmon (*e.g.*, from the river Neva) however, the impact is greatly reduced due to some form of innate resistance that regulates parasite numbers, resulting in fewer host mortalities. *Gyrodactylus salaris* is known from 17 European states; its status in a further 35 states remains unknown; the UK, the Republic of Ireland and certain watersheds in Finland are free of the parasite. Thus, the parasite poses a serious threat if it emerges in Atlantic salmon rearing regions throughout Europe. At present, infections are generally controlled via extreme measures such as the treatment of entire river catchments with the biocide rotenone, in order to remove all hosts, before restocking with the original genetic stock. The use of rotenone in this way in EU countries is unlikely as it would be in contravention of the Water Framework Directive. Not only are such treatments economically and environmentally costly, they also eradicate the potential for any host/parasite evolutionary process to occur. Based on previous studies, UK salmon stocks have been shown to be highly susceptible to infection, analogous to Norwegian stocks. The present study investigates the impact of a *G. salaris* outbreak within a naïve salmon population in order to determine long-term consequences of infection and the likelihood of coexistence. Simulation of the salmon/ *G. salaris* system was carried out via a deterministic mathematical modelling approach to examine the dynamics of host-pathogen interactions. Results indicated that in order for highly susceptible Atlantic strains to evolve a resistance, both a moderate-strong deceleratingly costly trade-off on birth rate and a lower overall cost of the immune response are required. The present study provides insights into the potential long term impact of *G. salaris* if introduced into *G. salaris*-free territories and suggests that in the absence of external controls salmon populations are likely to recover to high densities nearing 90% of that observed pre-infection.
**Introduction**

*Gyrodactylus salaris* Malmberg, 1957 is a viviparous (i.e., live-bearing) freshwater ectoparasite that infects both wild and farmed populations of Atlantic salmon (*Salmo salar* L.), potentially resulting in juvenile host mortality. It is an Office International des Epizooties (OIE) listed pathogen that was first described from the fins and skin of a Baltic Atlantic salmon strain from a hatchery in Sweden located near the Indalsälven river [1]. The parasite is believed to be native to the waters of northern Russia, western Sweden and northern Finland [2], but is now known to be widely distributed throughout Europe [3–10] and recently confirmed in Romania [11]. In Norway, the parasite has caused catastrophic damage to wild populations of Atlantic salmon parr since it was first observed in the mid-1970s after a period of mass salmon mortality [12–15]. Moreover, this parasite is known to have been introduced to Norway on at least three separate occasions [16] and can reduce salmon stock in rivers by approximately 85% on average [10]. Within 5 years of initial introduction to a susceptible host population reductions in outbound smolts can be as high as 98% [10,12,17]. This has caused severe damage to the Norwegian economy and to wild salmon fisheries. Although infections in salmon hatcheries have been reported, such infections are more readily controlled, however, if left untreated salmon mortality can reach 100% [10]. In the years post introduction, *G. salaris* has been reported from 50 rivers, 13 Atlantic salmon hatcheries and 26 rainbow trout (*Oncorhynchus mykiss* Walbaum) hatcheries in Norway and subsequently managed through coordinated intervention [18]. Subsequent losses to the Norwegian salmon industry up until 2004 exceeded US$ 655m [19]. The last time loss figures were estimated annual loss of wild juvenile salmon was suggested to be in the region of 250 - 500 metric tonnes as a consequence of parasitic infection reducing the average density of salmon parr in infected rivers [19]. Such annual loss costs the Norwegian economy over US$ 55m per annum through surveillance and eradication (circa US$ 23m per annum) along with losses incurred by fisheries, associated industries and
tourism (circa US$ 34m per annum) [14]. Hence, *G. salaris* poses a serious threat if it establishes in territories that are currently *G. salaris* free [9].

Though *G. salaris* has had a huge impact in Norway, some Baltic strains of Atlantic salmon appear to be more resistant to the parasite than the Atlantic strains [19]. Bakke *et al.* [20] was the first study to show a difference in the immune response between two strains of salmon. In particular, they showed that parasite numbers grew exponentially on individual fish from an Atlantic strain of Atlantic salmon from the rivers Lone and Alta (Norway), whereas on a Baltic strain of Atlantic salmon from the river Neva (Russia) there was some initial growth in parasite numbers, but those numbers peaked and then generally decreased to zero. This clearly demonstrated some differences in susceptibility of these salmon strains to *G. salaris* through the ability of the some Baltic strains to exhibit some form of resistance or immune response [19–23]. It has been highlighted that the resistance observed in some Baltic salmon strains, such as those from the Neva river, is due to the presence of the parasite in the Baltic watershed since the last glacial period allowing an evolutionary selection process within the host [22]. This supports the hypothesis that *G. salaris* is a recent (c. 40 years) introduction to Norwegian rivers and potentially explains why Norwegian Atlantic salmon are particularly susceptible to the parasite.

Due to the impact of *G. salaris* on Norwegian salmon, extreme measures have been taken to try and control and eradicate the parasite. These measures include the treatment of entire river catchments with the biocide rotenone [24] to remove all hosts (and hence, *G. salaris*), before restocking with the original genetic stock [12,14,25,26]. The use of rotenone in this way in EU countries is unlikely as it would be in breach of the Water Framework Directive [27]. Not only are such treatments economically and environmentally costly, they also eradicate the potential for any host/parasite evolutionary process to occur.
Currently the only European countries recognised as free from *G. salaris* infection are the United Kingdom [28,29], the Republic of Ireland [9,30,31], and some areas of Finland [9,32]. Other countries such as Portugal, Spain and France, where *G. salaris* has been previously recorded, are believed to be misidentifications with a morphologically similar species *Gyrodactylus teuchis* Lautraite, Blanc, Thiery, Daniel et Vigneulle, 1999 [32,33]. The collection of further material from these states is required to determine their current *G. salaris* status. Recently, however, it was proposed that *G. salaris* and *G. thymalli* Žitňan, 1960, another morphologically similar and closely-related, but benign parasite of grayling, *Thymallus thymallus* L., may represent a single species of *Gyrodactylus* that comprises several pathogenic and non-pathogenic strains on a number of primary hosts [34]. The study [34] analysed microRNA loci from a small number of populations of *Gyrodactylus* from Atlantic salmon and grayling hosts and made the proposal that the two species should be synonymised, however, this has not yet been formally accepted by the OIE and as such this synonymisation is yet to be accepted by the scientific community [11].

Despite the fact *G. salaris* is not present in the UK but *G. thymalli* is, it has been demonstrated that UK salmon populations have similar levels of susceptibility to infection as those in Norway [15,23,35,36]. Due to this, *G. salaris* is regarded to pose a serious disease threat to the UK’s valuable wild and farmed salmon populations [37]; a report to the Scottish Government advised if *G. salaris* were introduced into Scotland, as an example of potential impact, then the potential losses would be estimated at £44.8 million per annum to the Scottish economy, £34.5 million to Scottish household income each year and 1,996 full time equivalent jobs lost in Scottish employment [38]. It is also likely that *G. salaris*, if introduced, would spread within and between UK rivers before it is detected [2]. Due to this, contingency plans were drawn up setting out a series of actions to follow in the event of an outbreak [37]. Using mathematical modelling approaches based on the existing knowledge of *G. salaris*, the present study aims to simulate salmon/ *G. salaris* interaction dynamics in order to investigate the
potential for natural recovery of susceptible salmon populations post introduction of \textit{G. salaris} infection.

The majority of previous mathematical modelling work concerning the salmon/\textit{G. salaris} system has been centred on risk and statistical analysis highlighting areas such as routes of infection, transmission and risk of introduction \cite{2,39–43}. Some work has been carried out to study the effects of \textit{G. salaris} on different stages of the salmon life-cycle \cite{44} as well as the effect of other gyrodactylid species such as \textit{Gyrodactylus turnbulli} Harris, 1986 on guppies, \textit{Poecilia reticulata} Peters, 1859 \cite{45,46}. More recently stochastic models have become popular in studying \textit{G. salaris} infections in salmon and modelling techniques such as Leslie matrix population models and individual based models have also been employed \cite{47–49}. Though a great deal of effort has been placed on understanding the risks and routes by which the parasite may be introduced, little has been done to predict its long-term impact. Moreover, not much is known about what may happen should control efforts similar to those employed in Norway not be possible.

In the present study a series of host-macroparasite models are developed, first considering a single fish host and incorporating that into a population model. The effects that an increased immune response has on the host and parasite populations are analysed demonstrating the difference in susceptibility between a highly susceptible salmon strain and a resistant strain. Finally, some mutation and replacement is incorporated to determine how strong an immune response the hosts develop and what types of trade-offs and parameter values are required to allow a fully susceptible host to evolve into a primarily resistant host.

\textbf{Methods}
Individual fish model

To model parasite numbers on an individual host a deterministic ordinary differential equation (ODE) approach is taken. For the number of parasites, $P$, a simple exponential growth model is assumed, with replication rate $\mu$, death rate $\epsilon$ and dislodgement rate $\lambda$. In addition, we include an immune response, $I$, exhibited by the host which activates at rate $m$ as parasite numbers grow; this in turn increases the parasite death rate by a rate $pI$. Finally, the immune response decays at a continuous rate $\xi$. The equations for these are shown in equation (1) below:

$$
\frac{dP}{dt} = P(\mu - \epsilon - pI - \lambda)
$$

$$
\frac{dI}{dt} = mP - \xi I
$$

(1)

Full salmon population model

The individual fish host model was expanded by scaling up the equations in (1), to a population of hosts and parasites. Here the host population, $H$, is assumed to follow a logistic growth function, $a$ being the birth rate, $b$ the natural death rate and $s$ representing density-dependent competition, with an additional death rate dependent on parasite burden, $aM$. The equations for average parasite burden, $M = P/H$, or density of parasite per host (where $P$ is the total on-host parasite density), and immune response, $I$, are taken from equation (1), but expanded in that the parasite burden decreases due to deaths of the host due to infection, $a$, and birth of new (initially parasite-free) hosts. The on-host parasite distribution is assumed to follow a Poisson distribution across the host population, which is taken into account in the parasite-induced death rate, $a$. Both Poisson and negative binomial distributions were
considered with each giving similar results, the Poisson however, simplified the model significantly and thus was chosen. The off-host parasite density, $W$, is assumed to increase as the parasites leave the host (either by choice or host death) and decrease due to parasite death, $\sigma$, or parasite latching on to hosts at a rate $\beta$, which in turn increases parasite burden. It is important to note that actual parasite death rates are highly dependent on many factors such as environmental conditions (e.g. temperature), water quality, salinity, etc. [50,51]. In the present study, however, we consider a simplified worst case scenario such that we have a highly pathogenic strain of parasite and a highly susceptible Atlantic salmon strain.

The dynamics for the model take the form in (2). Further details of the model’s derivation are presented in the Supplementary Information (Appendix S1). Parameter values used in all models are given in Table 1. Parameter values regarding the UK were used where available.

\[
\begin{align*}
\frac{dH}{dt} &= (a - b - sH)H - aMH \\
\frac{dM}{dt} &= (\mu - \varepsilon - \rho l - \lambda - a - a - a)M + \beta W \\
\frac{dl}{dt} &= mM - \xi l \\
\frac{dW}{dt} &= MH[\lambda + b + sH + a(1 + M)] - \sigma W - \beta WH
\end{align*}
\]

(2)
### Table 1. List of parameter values used to inform salmon/G. salaris host parasite models.

| Parameter | Description                                      | Estimate/day | Source                                      |
|-----------|--------------------------------------------------|--------------|---------------------------------------------|
| \(a\)    | Maximum salmon birth rate                        | 0.02         | Assumed                                     |
| \(b\)    | Salmon natural death rate                        | 0.00057      | [52]                                        |
| \(K\)    | Salmon carrying capacity                         | 0.125        | [52]                                        |
| \(s\)    | Density dependent constraint                     | 0.000155     | Estimated using \(K\) for 1000 m²          |
| \(\mu\)  | \(G. salaris\) birth rate (Norway)              | 0.1825       | [20]                                        |
|           | \(G. salaris\) birth rate (UK)*                 | 0.1708       | [15]                                        |
| \(\epsilon\) | \(G. salaris\) on-host death rate              | 0.08         | [50]                                        |
| \(\sigma\) | \(G. salaris\) off-host death rate              | 0.14-0.17    | [42]                                        |
| \(\lambda\) | Rate the parasites leave the hosts              | 0.06         | Assumed                                     |
| \(\beta\) | Parasites attach rate to hosts                   | 0.0585       | Assumed                                     |
| \(\alpha\) | Parasite induced death rate of host              | 0.02         | [45]                                        |
| \(m\)    | Rate hosts develop an immune response            | 0 – 0.0175   | Assumed                                     |
| \(\xi\)  | Decay rate of immune response                    | 0.0055       | Assumed                                     |
| \(\rho\) | Rate of increase in parasite mortality due to resistance | 1            | Adjusted in values of \(m\)                |

* parameter value used in this study

With macro-parasite models, such as those used in the present study, fish-to-fish transmission is not shown explicitly in the model, but is rather an implicit feature modelled.
through the distribution of parasites across the fish population. This is due to the fact that \( P \) gives the total number of on-host parasites which remains unchanged as parasites switch between fish hosts, and due to the large number of parasites involved in these systems, the effect on the distribution of parasites is negligible.

**Results**

**Single host model**

Using the single host model, equation (1), two different cases were considered (Fig 1): firstly, a highly susceptible Atlantic salmon strain with no immunity, \( m \approx 0 \); secondly, a resistant salmon strain, \( m > 0 \) (\( m = 0.0175 \)). Model simulations showed parasite numbers grew exponentially on the susceptible host, whereas on the resistant host parasite numbers decayed to zero. In the case of the resistant host initial parasite growth over the first 7 days was similar to the highly susceptible host, however, parasite population growth slowed thereafter, peaking at around 20 days, before decreasing to zero/low levels. These behaviours approximately follow the experimental results observed by Bakke *et al.* [20] at water temperatures of 12°C on Atlantic Lone and Baltic Neva salmon hosts.

**Fig 1:** Output from the model in (1) for parasite numbers, with \( m = 0 \) (susceptible salmon strain – solid line) and \( m > 0 \) (resistant salmon strain – dashed line).
Full salmon population model

Firstly, the model in (2) was simulated to consider a fully susceptible host with a negligible immune response, \( i.e. m \approx 0 \). Here, following the introduction of the parasite into the system the model shows a fast drop in the number of hosts. This mirrors the results in the field, \( e.g., \) in Norway where the parasite can reduce the salmon parr population by up to 98% within 5 years [12]. As host extinction has not been witnessed, and the average reduction in salmon is 86% (and sometimes lower), we can assume that although \( m \neq 0 \), it must be very small. As we increase the amount of immune response, \( m \) (Fig 2A), we very quickly see that the host (equilibrium) population recovers and the average parasite burden decreases. In fact, only negligible values for \( m \) produces a reduction approaching 100%, and even a small amount of resistance significantly improves host population size. Moreover, host numbers approach their pre-infection levels, and parasite burden approaches zero, as \( m \) gets large. Interestingly the greatest effect on host and parasite numbers occurs at lower increases in immune response \( m \), with only marginal effects for larger \( m \).

Fig 2: Plot of host (equilibrium) population \( H \) (solid line) and parasite burden \( M \) (dashed line). (A) with no trade-off; (B) with a linear trade-off on host birth rate. The dotted line represents the (fully susceptible) host population before the parasite outbreak.

The trade-off

So far we have assumed that the immune response mounted by the host is cost free. This, however, has been shown not to be the case. One prime example of this is a study of furunculosis in brook trout, \( Salvelinus fontinalis \) (Mitchill) [53], in which it was shown that an increase in immunity had a negative effect on the host’s birth rate; they observed approximately a 7 to 12% decrease in the birth rate of the trout that exhibited resistance to infection. Although
there is no evidence to support or deny that a similar trade-off exists in salmon, for the remainder of this study we hypothesise there is a cost of the immune response. In particular, we take a trade-off such that the development of an effective immune response, as measured here by \( m \), can have a significant negative effect on host birth rate \( a \), such that \( a = a(m) \) with \( a'(m) < 0 \).

Although the form of \( a(m) \) is unknown, we make two assumptions: i) when \( m = 0 \), \( a = 0.02 \) (maximum birth rate) representing a highly susceptible salmon strain, and ii) when \( m = 0.0175 \) (our maximum resistance), birth rate \( a \) is reduced by 10% representing a resistant salmon strain. We initially take a linear trade-off (straight line) passing through these two points to allow us to interpolate \( a \) for intermediate \( m \).

The addition of this trade-off has a marked effect on the host population. In particular, at high levels of immune response, \( m \), the cost of a lower birth rate begins to outweigh the benefit of higher immune response (and subsequent lower parasite burden) and the host population begins to decrease (Fig 2B). Here an optimal level of immunity now exists which maximises the host population when \( m = 0.010 \).

**Mutation and replacement of hosts**

The optimal immune response observed may not, however, represent the level of \( m \) that the host species evolve to; this instead would likely be determined by the level of \( m \) which optimises the growth rate of the host population. To study the long-term evolution of immune response, we take a mutation and replacement approach, broadly following that of adaptive dynamics [54].

Consider a single resident host strain of salmon, with immune response \( m \) and population density \( H \) existing alone in an environment, with the dynamics as given in equation (2). Now suppose a mutation creates a host with slightly different immune response \( \hat{m} \), with population density \( \hat{H} \). Mutations are generally small, and hence, the difference between \( m \) and \( \hat{m} \) is small. Here \( \hat{M} \) and \( \hat{I} \) are the (average) parasite burden and immune response for this
mutant host strain. If this new type is initially rare, then we can write down the fitness of this mutant type, \textit{i.e.} the long-term growth rate of this mutant population, as

\begin{equation}
    r(\tilde{m}, m) = a(\tilde{m}) - b - sH(m) - a\tilde{M}(\tilde{m}, W(m))
\end{equation}

Here $\tilde{M}$ is the average parasite burden on a mutant host. We make the assumption that parasites will reach their "average" (equilibrium) burden on the new mutant host type $\tilde{M}$ quickly, when compared to the natural fish lifespan - a reasonable assumption given the much shorter generation time of the parasite. The full derivation of the fitness is given in the Supplementary Information (Appendix S3). If the fitness is positive, then the mutant host type will increase in number, generally replacing the existing resident host type, whereas if the fitness is negative the mutant will die out. For simplicity, we assume no 'intermediate strains' due to cross-breeding. The fitness is used to calculate the location of the evolutionary singular point and determine whether it is an evolutionary steady state, ESS, \textit{i.e.} an evolutionary end point.

To demonstrate the evolutionary behaviour more clearly, we numerically simulate evolution using a similar mutation and replacement approach, using the full mutant-resident dynamics – details of which are presented in the Supplementary Information (Appendix S3). This has been shown to be a good approximation to the analytical approach using the fitness in (3) and has the benefit of not making the assumption about the parasite burden being at equilibrium. Starting from a highly susceptible salmon strain, we plot how $m$ evolves through time. Fig 3A plots the strains present following each mutation and shows how $m$ evolves over time with a (linear) trade-off. Here 'time' means the number of mutation events that occur – as we do not currently know how often mutations occur, we leave time deliberately in terms of these mutation events. In addition, the colouring represents the total host population present.
For the first 100 time steps, the system is parasite-free, hence minimal resistance and maximum host population (Fig 3A). At time step 100, however, we introduce a small number of (free-living) parasites. Immediately the population of host drops (Fig 3B). Resistance then begins to be selected for, leading to an increase in $m$ (Fig 3A). This in turn leads to an increase in host population and a lower parasite burden (Fig 3B). The level of immune response eventually settles at an intermediate level, i.e. an ESS, with the host population normally distributed about this resistance level (Fig 3A- inset). This is at approx. $m = 0.0075$ here, slightly below the optimal $m (\approx 0.010)$ which maximises the host population.

**Fig 3:** In (A) we plot how $m$ evolves over time, with a linear trade-off; the colour of the line denotes the total host population at that time. The inset graphs give the distribution of resistance levels in the host population at time=100, just prior to parasite invasion, and at time=300, when the population reaches its ESS. In (B) we plot the host population and parasite burden over time, corresponding to $m$ evolving.

**Trade-off shape**

So far we have only considered a linear trade-off - whereby each benefit (i.e. unit increase in immune response, $m$) always comes at the same cost (i.e. same decrease in birth rate, $a$). We now vary the trade-off shape by means of a parameter $\theta$ (see Supplementary Information, Appendix S3, for specific details). Specifically, a positive $\theta$ represents an 'acceleratingly costly trade-off', whereby each benefit comes at an increasing (accelerating) cost (i.e. larger decrease in birth rate, $a$); with larger $\theta$ giving a greater effect. Conversely, a negative $\theta$ represents a 'deceleratingly costly trade-off', whereby each benefit comes at a decreasing (decelerating) cost (i.e. a smaller decrease in birth rate, $a$). Finally $\theta = 0$ represents a linear trade-off [54].
In Fig 4 we plot how the evolutionary singular point (ESS) \( m^* \) changes as we change the shape of the trade-off (\( \theta \) values); where the ESS is denoted by the thick black line. The host evolves to increase their resistance level \( m \) if currently below the ESS, and evolve to decrease resistance if above. In addition, the contour lines represent the equilibrium host density. We immediately gain two main results from this. Firstly, that the evolutionary singular points are always just below the maximum host density for each specific value of \( \theta \), meaning that the optimal value of \( m \) which maximises the host population is not the same value of \( m \) that maximises host fitness. Secondly, for strong deceleratingly costly trade-offs, as \( \theta \to -1 \), the host evolves to maximise the immune response \( m \), whereas for weakly deceleratingly costly or acceleratingly costly trade-offs, the host evolves to an intermediate value of \( m \). This suggests a limited range of trade-offs that allow a highly susceptible salmon host strain to evolve into a highly resistant host strain.

Fig 4: Plot of the evolutionary singular point (ESS - thick black line) for various shapes of trade-off. Here \( \theta < 0 \) represents a deceleratingly costly trade-off; \( \theta > 0 \) represents an acceleratingly costly trade-off; and \( \theta = 0 \) (dashed line) represents a linear trade-off – as taken in Fig 3 simulation. The host evolves such that the immune response \( m \) either increases or decreases (vertically on the plot) to the singular point – see Supplementary Information (Appendix S4) for derivation of this line. The thin contour lines represent the total host population size for corresponding values of \( m \) and \( \theta \). The parameters are as given in Table 1.

Virulence

In Fig 5A, we plot the evolutionary singular points (ESS) for varying levels of parasite virulence, in terms of a higher or lower parasite-induced host death rate, \( \alpha \). Higher levels of virulence, common in \( G \) salaris [15,23,35,36], encourages the evolution of a stronger immune response.
Cost of resistance

In Fig 5B, we show the equivalent results for the lower and upper estimates for the cost of resistance, as given by Cipriano et al. [53], 7% and 12% respectively (as opposed to the ‘averaged’ 10% initially taken). As would be expected, the location of the evolutionary singular points (ESS) is lowered as the cost of resistance is increased, implying that the hosts evolve a lower immune response, $m$, if more costly. This suggests that for the host to evolve into a highly resistant strain, the cost of being highly resistant must not be too high.

Fig 5: Plot of the evolutionary singular point (ESS) for various shapes of trade-off: $\theta < 0$ deceleratingly costly, $\theta > 0$ acceleratingly costly and $\theta = 0$ (dashed line) linear. The colour of each line is defined by the average host density along that line, as represented on the colour bar. In (A) the virulence of the pathogen is varied, with $\alpha = 0.02$ being the baseline value. In (B) the cost of resistance is varied, with 10% being the baseline.

Discussion

Wild Atlantic salmon populations the world over are currently threatened, with numbers in some regions in decline [55]. The catastrophic impact that infections by *G. salaris* can have on susceptible salmon populations, and the consequential financial implications, have already been witnessed in Norway [12,17,56,57]. In the years post introduction to Norway, *G. salaris* has since been reported from many other river systems throughout Europe [3–8,10]. The aim of the present study was to explore the long-term interactions between populations of Atlantic salmon and the monogenean parasite *G. salaris* in order to make predictions on the natural recovery of salmon populations post introducing such an infection into an environment containing susceptible salmon host populations such as the United Kingdom.
In the present study models were used to study the possible differences between strains of Atlantic salmon to determine the mechanisms evolved by some Baltic strains in order to be able to beat infection and in some cases coexist with low levels of *G. salaris* infection. Model outcomes have highlighted that simple host-parasite models can show the varying levels of resistance as seen in the Atlantic Lone and Baltic Neva salmon systems, with the addition of an immune response. Models were used to investigate the possibility highly susceptible strains of Atlantic salmon evolving traits and resulting trade-offs to become more like their resistant counterparts.

Results from the present study highlight salmon will evolve to a more resistant state and therefore be able to naturally recover from *G. salaris* infection if the salmon immune response is allowed to evolve. This evolution would be subject to a trade-off such that host birth rate is negatively correlated with resistance. Such recovery would result in host coexistence, potentially at relatively high host densities, nearing 90% to that observed in the absence of infection, with low parasite densities. The level of immune response however depends on several factors: In order for a susceptible host to gain the level of resistance witnessed in some Baltic salmon strains, it requires both a moderate-strong deceleratingly costly trade-off (*i.e.*, the host pays a large cost in the creation of the immune response, for low $m$, and then the additional costs for improving that immune response, increasing $m$, are less and reducing) and a lower overall cost of the immune response. In addition, the virulence of parasite can play a significant part, with higher virulence rates leading to lower host population sizes but higher resistance levels; conversely, lower virulence rates leads to higher host populations with lower resistance levels. For this reason, the water chemistry can play a crucial part in how salmon evolve as identical strains of parasite can have different virulence rates solely due to environmental factors.

In general, mathematical models represent a simplified version of a system, as such, there are always going to be certain limitation. Future studies would do well to build on the models herein and explicitly model the seasonal effects and implications of the salmon and
gyrodactylid life-cycles. Salmon spawning, for example, primarily takes place once a year between mid-October and late February [58]. Similarly, salmon do not spend their entire life in a river and in fact spend the majority of their adult life at sea, returning to their natal river to spawn. Though it is possible for some salmon parr to mature sexually in a river without the need to run to sea, and hence, stay to participate in spawning [59]. Such behaviours will have an important impact on the length of time it would take for a population of salmon to recover from *G. salaris* infections due to the time between salmon leaving and returning to infected rivers.

Salinity and water temperature are very important in determining *G. salaris* survival. *Gyrodactylus salaris* is a freshwater parasite and survival is only possible in waters with a salinity between 0 – 20ppt at temperatures of 3°C - 20°C [50,51]. The survival of *G. salaris* in low salinity waters has been shown to be negatively correlated with water temperature and hence, parasites can survive longer, both on and off a host, in such waters at lower temperatures [51]. Environment can also play an important role; in situations where water velocity is high, detached parasites have the potential to drift further down a river and infect new populations of hosts. Infection may also have an impact on the way in which salmon interact with each other, for example, in populations of guppies, *P. reticulata*, (where individuals are infected with *Gyrodactylus turnbulli*) females have been observed preferring, and selecting, males with low parasite burdens [60]. Furthermore, changes in host feeding behaviour has also been witnessed with feeding response and feeding activity significantly negatively correlated with parasite load [61].

Whilst the varying degrees of pathogenicity of the different *G. salaris* strains was not explicitly modelled in the present study, future studies would do well to include such information into predictive models. Different strains of *G. salaris* have been shown to have varying effects on salmon hosts [16]. The three currently known clades of *G. salaris* include *G. salaris sensu stricto* - a highly pathogenic strain only found on Atlantic salmon (Clade I); a strain found on salmon from the river Göta älv in Sweden (Clade 2); and a strain that was found on salmon from the rivers Lærdalselva, Drammenselva and Lierelva in Norway and on rainbow
trout from a fish farm in Lake Bullaren, Sweden [16]. A further strain of *G. salaris* has been found on rainbow trout in Denmark [3,4]. This variant of the *G. salaris* parasite shows low virulence towards Atlantic salmon and under experimental conditions, on isolated hosts, this strain showed limited reproduction or no establishment at all [62]. Lindenstrom et al. [63], however, observed high susceptibility to this strain in rainbow trout and noted that this strain of the parasite greatly resembles *G. salaris sensu stricto*.

As highlighted earlier, fish-to-fish transmission was modelled through the distribution of parasites across the fish population and not as an explicit feature in the model. The models proposed consider the total densities of a *G. salaris* population within a salmon host population. It would also be interesting to take an approach looking into the density of *G. salaris* populations on individual hosts within a population with particular focus on the impact that fish-to-fish transmission has on the dynamics of infection. It is known that juvenile Atlantic salmon are highly territorial [59,63] and hence have a high chance of becoming infected due to fish-to-fish contact when defending a territory against an infected individual. Moreover, fish-to-fish contact between dead infected hosts and live uninfected hosts as well as live infected hosts and live uninfected hosts also provide important routes for *G. salaris* spread [64,65].

Aggregation of parasites on hosts also has an important impact on the evolutionary and population dynamics of both parasites and hosts [66,67]. Many studies have been carried out in this area in order to develop our understanding of what causes heterogeneity in the distribution of macroparasites within a host population [68]. Parasite aggregation in the wild is often complex, in macro-parasitic infections the majority of hosts are observed harbouring a low number of parasites with a minority of hosts harbouring a large number [69]. Such skewed aggregations have been shown to follow a negative binomial distribution [66,67,69]. The negative binomial distribution, (defined as $s^2 = m + m^2/k$, where $s^2$ and $m$ are the variance and mean respectively) quantifies the (inverse) degree of aggregation via the parameter $k$ [70] such that for small $k$ parasite aggregation is increased, whereas for large $k$ aggregation decreases. The negative binomial distribution converges on the logarithmic series as $k \to 0$ and on the
Poisson for $k \geq 20$ [68,71]. Due to the complicated life-cycle of *G. salaris* and its similarities with micro- as well as macro-parasites we used a Poisson (defined as $s^2 = m$) to model parasite aggregation in the present study, thus, allowing parasites to be randomly (and evenly) distributed throughout the host population. This simplified model analyses considerably whilst still allowing for important observations to made on the dynamics of infection. Previous studies have considered a Poisson distribution when modelling free-living *G. salaris* parasites [72]. Moreover, the effect on the distribution of parasites is negligible due to the large number of parasites considered in the present models.

Even though the literature concerning *G. salaris* infections in salmon is vast, models would greatly benefit from more accurate and up to date parameter estimates. Experimental studies undertaken exclusively for this reason would be worthwhile in order to obtain estimates for currently unknown parameters. Through our research we have determined that more data are required in order to accurately parameterise the rate at which parasites leave, attach to and kill hosts.

At present the United Kingdom and Ireland are the only known countries to officially establish complete freedom from *G. salaris* infections [10,28–30,37]. As highlighted earlier, Atlantic salmon populations in the UK are believed to be just as susceptible as those found in Norway [15,35], hence, if *G. salaris* was introduced a similar environmental impact to that of Norway can be expected. Extreme measures have been adopted in an attempt to control and eradicate *G. salaris* infections. While eradication is preferred, this rarely happens and hence “management and control” is what is actually being carried out and alternative methods of treatment such as aluminium have been trialled [73]. It is understandable that survivors are undesirable as we may see the development of resistance in the parasite population with consequentially continued catastrophic effects on the host population, however, we also would like to see the evolutionary process occur where there is adaptation or co-evolution to the extent that parasite and host to co-exist without mortality and parasite numbers are maintained
at low levels or are removed by the host. Our results highlight that the current practice of
treating entire river catchments with rotenone before restocking with salmon from the original
genetic stock [12,14,25,26] may be severely damaging the potential for any evolutionary
process to occur.

Results from the present study have provided evidence that in the absence of
intervention salmon populations should naturally recover from *G. salaris* infection, however, the
timescale required for this to happen remains unknown. Furthermore, model output suggests
susceptible populations would evolve such that they reach a level of resistance required to
coexist with the parasite and recover to relatively high densities, nearing 90% of that observed
pre-infection. *Gyrodactylus salaris* and its impact on susceptible hosts must continue to be
studied in order to aid in contingency planning and defence against introduction and
emergence.
References

1. Malmberg G. On the occurrence of *Gyrodactylus* on Swedish fishes. Skr utgivna av Södra Sveriges Fisk Årsskrift. 1957;1956:19–76.

2. Peeler EJ, Thrush MA. Qualitative analysis of the risk of introducing *Gyrodactylus salaris* into the United Kingdom. Dis Aquat Organ. 2004;62:103–13.

3. Buchmann K, Bresciani J. Parasitic infections in pond-reared rainbow trout *Oncorhynchus mykiss* in Denmark. Dis Aquat Organ. 1997;28:125–38.

4. Nielsen CV, Buchmann K. Occurrence of *Gyrodactylus* parasites in Danish fish farms. Bull Eur Assoc Fish Pathol. 2001;21:19–25.

5. Cunningham CO, Collins CM, Malmberg G, Mo TA. Analysis of ribosomal RNA intergenic spacer (IGS) sequences in species and populations of *Gyrodactylus* (Platyhelminthes: Monogenea) from salmonid fish in northern Europe. Dis Aquat Organ. 2003;57:237–46.

6. Ziętara MS, Rokicka M, Stojanovski S, Skorkowski EF, Lumme J. Alien mitochondrial DNA in variant clones of *Gyrodactylus salaris* indicates a complex hybrid history in salmonid farms. Parasitologia. 2007;49:119.

7. Rokicka M, Lumme J, Ziętara MS. Identification of *Gyrodactylus* ectoparasites in Polish salmonid farms by PCR-RFLP of the nuclear ITS segment of ribosomal DNA (Monogenea, Gyrodactylidae). Acta Parasitol. 2007;52(3):185–95.

8. Paladini G, Gustinelli A, Fioravanti ML, Hansen H, Shinn AP. The first report of *Gyrodactylus salaris* Malmberg, 1957 (Platyhelminthes, Monogenea) on Italian cultured stocks of rainbow trout (*Oncorhynchus mykiss* Walbaum). Vet Parasitol. 2009;165(3):290–7.

9. Paladini G. Aspects of systematics and host specificity for *Gyrodactylus* species in aquaculture. University of Stirling; 2012.

10. OIE. Chapter 2.3.3. Infection with *Gyrodactylus salaris*. In: Manual of Diagnostic Tests for Aquatic Animals 2016. Paris: Office International des Epizooties; 2016. p. 499–510.

11. Hansen H, Cojocaru C-D, Mo TA. Infections with *Gyrodactylus* spp. (Monogenea) in Romanian fish farms: *Gyrodactylus salaris* Malmberg, 1957 extends its range. Parasit Vectors. 2016;9(1):444.

12. Johnsen BO, Jensen AJ. The *Gyrodactylus* story in Norway. Aquaculture. 1991;98:289–302.

13. Johnsen BO, Møkkelgjerd PI, Jensen AJ. The parasite *Gyrodactylus salaris* on salmon parr in Norwegian rivers, status report at the beginning of year 2000. NINA Oppdragsmeld. 1999;617:1–129.

14. Bakke TA, Cable J, Harris PD. The biology of gyrodactylid monogeneans: The “Russian-doll killers.” Adv Parasitol. 2007;64:161–460.
15. Paladini G, Hansen H, Williams CF, Taylor NG, Rubio- Mejía OL, Denholm SJ, et al. Reservoir hosts for *Gyrodactylus salaris* may play a more significant role in epidemics than previously thought. Parasit Vectors. 2014 Dec 20;7(1):576.

16. Hansen H, Bachmann L, Bakke TA. Mitochondrial DNA variation of *Gyrodactylus* spp. (Monogenea, Gyrodactylidae) populations infecting Atlantic salmon, grayling, and rainbow trout in Norway and Sweden. Int J Parasitol. 2003;33:1471–8.

17. Mo TA. Status of *Gyrodactylus salaris* problems and research in Norway. In: Parasitic Diseases of Fish. Pike A, Lewis J, editors. Dyfed, Wales, UK: Samara Publishing Ltd; 1994. 43-58 p.

18. Hytterød S, Linaker ML, Hansen H, Mo TA, Tavornpanich S. The surveillance and control programme for *Gyrodactylus salaris* in Atlantic salmon and rainbow trout in Norway 2014. In: Surveillance programmes for terrestrial and aquatic animals in Norway Annual report 2013. Oslo: Norwegian Veterinary Institute; 2015.

19. Bakke TA, Harris PD, Hansen H, Cable J, Hansen LP. Susceptibility of Baltic and East Atlantic salmon *Salmo salar* stocks to *Gyrodactylus salaris* (Monogenea). Dis Aquat Organ. 2004;58:171–7.

20. Bakke TA, Jansen PA, Hansen LP. Differences in the host resistance of Atlantic salmon, *Salmo salar* L, stocks to the monogenean *Gyrodactylus salaris* Malmberg, 1957. J Fish Biol. 1990;37:577–87.

21. Cable J, Harris PD, Bakke TA. Population growth of *Gyrodactylus salaris* (Monogenea) on Norwegian and Baltic Atlantic salmon (*Salmo salar*) stocks. Parasitology. 2000;121:621–9.

22. Bakke TA, Harris PD, Cable J. Host specificity dynamics: observations on gyrodictylid monogeneans. Int J Parasitol. 2002;32:281–308.

23. Dalgaard MB, Nielsen C V., Buchmann K. Comparative susceptibility of two races of *Salmo salar* (Baltic Lule river and Atlantic Conon river strains) to infection with *Gyrodactylus salaris*. Dis Aquat Organ. 2003;53:173–6.

24. IPCS-INCHEM. International Programme on Chemical Safety - Rotenone: health and safety guide. Geneva (Health and safety guide; No. 73). ISBN 9241510730. ISSN 0259-7268. 1992.

25. Holm A, Molander P, Lundanes E, Greibrokk T. Determination of rotenone in river water utilizing packed capillary column switching liquid chromatography with UV and time-of-flight mass spectrometric detection. J Chromatogr A. 2003;983:43–50.

26. Salte R, Bentsen HB, Moen T, Tripathy S, Bakke TA, Ødegård J, et al. Prospects for a genetic management strategy to control *Gyrodactylus salaris* infection in wild Atlantic salmon (*Salmo salar*) stocks. Can J Fish Aquat Sci. 2010;67:121–9.

27. European Parliament Council. Directive 2000/60/EC of the European Parliament and of the Council of 23 October 2000 establishing a framework for Community action in the field of water policy. 2000;
28. European Commission. COMMISSION DECISION of 29 April 2004 implementing Council Directive 91/67/EEC as regards measures against certain diseases in aquaculture animals. Official Journal of the European Communities. 2004.

29. European Commission. COMMISSION DECISION of 5 April 2006 amending Decision 2004/453/EC as regards Sweden and the United Kingdom. Official Journal of the European Communities. 2006.

30. Shinn AP, Sommerville C, Gibson DI. Distribution and characterization of species of *Gyrodactylus* Nordmann, 1832 (Monogenea) parasitizing salmonids in the UK, and their discrimination from *Gyrodactylus salaris* Malmberg, 1957. *J Nat Hist.* 1995;29:1383–402.

31. Platten M, McLoughlin M, Shinn AP. Distribution and identification of gyrodactyloid species in fish farms and rivers of Northern Ireland. Vet Rec. BMJ Publishing Group Limited; 1994 Oct 22;135(17):411–2.

32. Lautraite A, Blanc G, Thiery R, Daniel P, Vigneulle M. Gyrodactyloids parasitizing salmonids in Brittany and Western Pyrenees water basins: Epidemiological features of infection and species composition. Bull Français la Pêche et la Piscic. 1999;355:305–25.

33. Cunningham CO, Mo TA, Collins CM, Buchmann K, Thiery R, Blanc G, et al. Redescription of *Gyrodactylus teuchis* Lautraite, Blanc, Thiery, Daniel & Vigneulle, 1999 (Monogenea: Gyrodactylidae); a species identified by ribosomal RNA sequence. *Syst Parasitol.* 2001;48(2):141–50.

34. Fromm B, Burow S, Hahn C, Bachmann L. MicroRNA loci support conspecificity of *Gyrodactylus salaris* and *Gyrodactylus thymalli* (Platyhelminthes: Monogenea). *Int J Parasitol.* 2014;44(11):787–93.

35. Bakke TA, MacKenzie K. Comparative susceptibility of native Scottish and Norwegian stocks of Atlantic salmon, *Salmo salar* L., to *Gyrodactylus salaris* Malmberg: Laboratory experiments. *Fish Res.* 1993;17:69–85.

36. Dalgaard MB, Larsen TB, Jørndrup S, Buchmann K. Differing resistance of Atlantic salmon strains and rainbow trout to *Gyrodactylus salaris* infection. *J Aquat Anim Health.* 2004;16:109–15.

37. Defra. Contingency Plan for Combating *Gyrodactylus salaris* in England. Available at www.defra.gov.uk. 2008.

38. Riddington G, Radford A, Paffrath S, Bostock J, Shinn A. An Economic Evaluation of the Impact of the Salmon Parasite *Gyrodactylus salaris* (Gs) Should it be Introduced into Scotland: Summary Report. Prepared for the Scottish Executive Environment and Rural Affairs Department, Project Number SAQ/001/05. 2006.

39. Paisley LG, Karlsen E, Jarp J, Mo TA. A Monte Carlo simulation model for assessing the risk of introduction of *Gyrodactylus salaris* to the Tana river, Norway. *Dis Aquat Organ.* 1999;37:145–52.

40. Høgåsen HR, Brun E. Risk of inter-river transmission of *Gyrodactylus salaris* by migrating Atlantic salmon smolts, estimated by Monte Carlo simulation. *Dis Aquat Organ.* 2003;57:247–54.
41. Peeler EJ, Gardiner R, Thrush MA. Qualitative risk assessment of routes of transmission of the exotic fish parasite *Gyrodactylus salaris* between river catchments in England and Wales. *Prev Vet Med.* 2004;64:175–89.

42. Peeler EJ, Thrush M, Paisley L, Rodgers C. An assessment of the risk of spreading the fish parasite *Gyrodactylus salaris* to uninfected territories in the European Union with the movement of live Atlantic salmon (*Salmo salar*) from coastal waters. *Aquaculture.* 2006;258:187–97.

43. Jansen PA, Matthews L, Toft N. Geographic risk factors for inter-river dispersal of *Gyrodactylus salaris* in fjord systems in Norway. *Dis Aquat Organ.* 2007;2:139–49.

44. Des Clercs S. Modelling the impact of disease-induced mortality on the population size of wild salmonids. *Fish Res.* 1993;17:237–48.

45. Scott ME, Anderson RM. The population dynamics of *Gyrodactylus bullatarudis* (Monogenea) within laboratory populations of the fish host *Poecilia reticulata.* *Parasitology.* 1984;89:159.

46. Scott ME. Experimental epidemiology of *Gyrodactylus bullatarudis* (Monogenea) on guppies (*Poecilia reticulata*): short and long-term studies. *Ecol Genet Host-Parasite Interact.* 1985;21–38.

47. Van Oosterhout C, Potter R, Wright H, Cable J. Gyroscope: An individual-based computer model to forecast gyrodactylid infections on fish hosts. *Int J Parasitol.* 2008;38:541–8.

48. Ramirez R, Harris PD, Bakke TA. An agent-based modelling approach to estimate error in gyrodactylid population growth. *Int J Parasitol.* 2012;42:809–17.

49. Denholm SJ, Norman RA, Hoyle AS, Shinn AP, Taylor NGH. Reproductive Trade-Offs May Moderate the Impact of *Gyrodactylus salaris* in Warmer Climates. *PLoS One.* 2013;8(10):e78909.

50. Jansen PA, Bakke TA. Temperature-dependant reproduction and survival of *Gyrodactylus salaris* Malmberg, 1957 (Platyhelminthes - Monogenea) on Atlantic salmon (*Salmo salar* L.). *Parasitology.* 1991;102:105–12.

51. Soleng A, Bakke TA, Hansen LP. Potential for dispersal of *Gyrodactylus salaris* (Platyhelminthes, Monogenea) by sea-running stages of the Atlantic salmon (*Salmo salar*): field and laboratory studies. *Can J Fish Aquat Sci.* 1998;55:507–14.

52. Hedger RD, Sundt-Hansen LE, Forseth T, Diserud OH, Ugedal O, Finstad AG. Modelling the complete life-cycle of Atlantic salmon (*Salmo salar* L.) using a spatially explicit individual-based approach. *Ecol Modell.* 2013;248:119–29.

53. Cipriano RC, Marchant D, Jones TE, Schachte JH. Practical application of disease resistance: a brook trout fishery selected for resistance to furunculosis. *Aquaculture.* 2002;206:1–17.

54. Geritz SAH, Kisdi É, Meszána G, Metz JAJ. Evolutionarily singular strategies and the adaptive growth and branching of the evolutionary tree. *Evol Ecol.* 1998;12(1):35–57.

55. WWF. The status of wild Atlantic salmon: a river by river assessment. 2001. 174 p.
56. Johnsen BO. The effect of an attack by the parasite *Gyrodactylus salaris* on the population of salmon parr in the river Lakselva, Misløv in northern Norway. *Astarte*. 1978;11:7–9.

57. Johnsen BO, Jensen AJ. Infestations of Atlantic salmon, *Salmo salar*, by *Gyrodactylus salaris* in Norwegian rivers. *J Fish Biol*. 1986;29:233–41.

58. Shearer WM. The Atlantic salmon: natural history, exploitation and future management. Oxford: Fishing News Books; 1992.

59. Crisp DT. Trout and Salmon: Ecology, Conservation and Rehabilitation. London: Blackwell Science Ltd; 2000. 324 p.

60. Kennedy CEJ, Endler JA, Poynton SL, H M. Parasite load predicts mate choice in guppies. *Behav Ecol Sociobiol*. 1987;21:291–5.

61. Van Oosterhout C, Harris PD, Cable J. Marked variation in parasite resistance between two wild populations of the Trinidadian guppy, *Poecilia reticulata* (Pisces: Poeciliidae). *Biol J Linn Soc*. 2003;79:645–51.

62. Lindenstrom T, Collins CM, Bresciani J, Cunningham CO, Buchmann K. Characterization of a *Gyrodactylus salaris* variant: infection biology, morphology and molecular genetics. *Parasitology*. 2003;127:165–77.

63. Keenlyside MHA, Yamamoto FT. Territorial behaviour of juvenile Atlantic salmon (*Salmo salar* L.). *Behaviour*. 1962;19:139–69.

64. Bakke TA, Harris PD, Jansen PA. The susceptibility of *Salvelinus fontinalis* (Mitchill) to *Gyrodactylus salaris* Malmberg (Platyhelminthes; Monogenea) under experimental conditions. *J Fish Biol*. 1992;41:499–507.

65. Soleng A, Jansen PA, Bakke TA. Transmission of the monogenean *Gyrodactylus salaris*. *Folia Parasitol* (Praha). 1999;46:179–84.

66. Anderson RM, May RM. Regulation and stability of host-parasite population interactions. I: Regulatory processes. *J Anim Ecol*. 1978;47:219–47.

67. May RM, Anderson RM. Regulation and stability of host-parasite population interactions. II: Destabilizing processes. *J Anim Ecol*. 1978;47:249–67.

68. Wilson K, Bjørnstad ON, Dobson AP, Merler S, Poglayen G, Read a F, et al. Heterogeneities in macro-parasite infections: patterns and processes. In: Hudson P, Rizzoli A, Grenfell B, Heesterbeek H, Dobson A, editors. The Ecology of Wildlife Diseases. Oxford: Oxford University Press; 2002.

69. Shaw D, Dobson A. Patterns of macroparasites abundance and aggregation in wildlife populations: a quantitative review. *Parasitology*. 1995;111(1995):111–33.

70. Bliss CI, Fisher RA. Fitting the negative binomial distribution to biological data. *Biometrics*. 1953;9(2):176–200.

71. Fisher RA, Corbet S, Williams CB. The Relation Between the Number of Species and the Number of Individuals in a Random Sample of an Animal Population. *J Anim Ecol*. 1943;12(1):42–58.
Høgåsen HR, Brun E, Jansen P a. Quantification of free-living Gyrodactylus salaris in an infested river and consequences for inter-river dispersal. Dis Aquat Organ. 2009;87(3):217–23.

Soleng A, Poleo ABS, Alstad NEW, Bakke TA. Aqueous aluminium eliminates Gyrodactylus salaris (Platyhelminthes, Monogenea) infections in Atlantic salmon. Parasitology. 1999;119:19–25.
Supporting Information

S1 Fig. Schematic representation of salmon-Gs model

S2 Fig. The trade-off between host birth rate, \( a \), and the rate hosts mount an immune response to the parasite (resistance), \( m \).

S1 File. Appendix S1 – Derivation of the model

S2 File. Appendix S2 – Equilibrium and stability analysis of model

S3 File. Appendix S3 – Derivation of the fitness of mutant type and trade-off

S4 File. Appendix S4 – Details of evolutionary simulations

S5 File. Appendix S5 - Supporting Information reference list.