Detailed Assessment of the Effects of Meteorological Conditions on PM$_{10}$ Concentrations in the Northeastern Part of the Czech Republic

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Abstract: This article assessed the links between PM$_{10}$ pollution and meteorological conditions over the Czech-Polish border area at the Trinec-Kosmos and Věřnovice sites often burdened with high air pollution covering the years 2016–2019. For this purpose, the results of the measurements of special systems (ceilometers) that monitor the atmospheric boundary layer were used in the analysis. Meteorological conditions, including the mixing layer height (MLH), undoubtedly influence the air pollution level. Combinations of meteorological conditions and their influence on PM$_{10}$ concentrations also vary, depending on the pollution sources of a certain area and the geographical conditions of the monitoring site. Generally, the worst dispersion conditions for the PM$_{10}$ air pollution level occur at low air temperatures, low wind speed, and low height of the mixing layer along with a wind direction from areas with a higher accumulation of pollution sources. The average PM$_{10}$ concentrations at temperatures below 1 °C reach the highest values on the occurrence of a mixing layer height of up to 400 m at both sites. The influence of a rising height of the mixing layer at temperatures below 1 °C on the average PM$_{10}$ concentrations at Trinec-Kosmos site is not as significant as in the case of Věřnovice, where a difference of several tens of µg·m$^{-3}$ in the average PM$_{10}$ concentrations was observed between levels of up to 200 m and levels of 200–300 m. The average PM$_{10}$ hourly concentrations at Trinec-Kosmos were the highest at wind speeds of up to 0.5 m·s$^{-1}$, at MLH levels of up to almost 600 m; at Věřnovice, the influence of wind speeds of up to 2 m·s$^{-1}$ was detected. Despite the fact that the most frequent PM$_{10}$ contributions come to the Trinec-Kosmos site from the SE direction, the average maximum concentration contributions come from the W–N sectors at low wind speeds and MLHs of up to 400 m. In Věřnovice, regardless of the prevailing SW wind direction, sources in the NE–E sector from the site have a crucial influence on the air pollution level caused by PM$_{10}$.

Keywords: mixing layer height; ceilometer; suspended particulate matter; air pollution; Czech-Polish border

1. Introduction

PM$_{10}$ is a problematic pollutant with a wide spectrum of effects on human health, mostly on the respiratory and cardiovascular systems of the human body. The hazards of this pollutant lie not only in its quantity, thus in high measured concentrations, but also in the morphology of the particles and their qualitative composition. The suspended particulate matter is capable of creating bonds, which may, of course, result in transferring a range of other elements, e.g., heavy metals and polycyclic aromatic hydrocarbons, into the human body. The effects on the human body have clinically proven to be negative, especially in the form of carcinogenesis [1–5], for some of these transported substances.
As for the areas of concern, the areas in Europe most polluted by suspended particulate matter are the Po Valley in northern Italy, the countries of Eastern Europe, including the Balkan Peninsula, and part of Central Europe. The southern rim of the area (of Central Europe) occupies the northeastern part of the Czech Republic, and spreads northwards (in terms of area) to the larger territory of Poland [6,7]. This Czech-Polish region is part of the Upper Silesian Basin, which is a black coal basin of European significance. The first (primitive) black coal mining in the Upper Silesian Basin began as early as 1657 on Polish territory south of Katowice; real industrial mining development started in the Czech part of the basin no earlier than the second half of the 19th century [8]. The mining industry became the basis for the development of other industrial branches. In the entire region of the Upper Silesian Basin, there is an intricately interconnected system of industry, traffic infrastructure, and high density of population with individual solid fuel heating, which is historically related to black coal burning. All of the transborder region of the Silesian Voivodeship in Poland and the Moravian-Silesian Region in the Czech Republic rank among the most urbanized and industrialized regions in Europe [9,10]. All these anthropogenic activities negatively influence all the environmental components, and pronounced changes in these components have been recorded since the 1970s. From the point of view of air quality improvement, there were significant positive changes in the 1990s during the economic restructuring of both post-communist countries (Poland and the then-Czechoslovakia). Another major step forward was the accession of both countries to the European Union and the necessity of adapting their legislation to the EU requirements [11,12]. Owing to the mentioned measures, there has been a decrease in the pollution load of the Czech-Polish region; the problems, however, remain, and the exceedance of the pollution limits is still frequent [13,14]. This especially concerns the suspended particulate matter and benzo(a)pyrene [15].

The particular air pollution level in the region of concern does not depend solely on the pollution source characteristics or the amount of discharged pollutants, but also on the physical geography and meteorological conditions of the area in question. It is the meteorological conditions themselves that determine the intensity and manner of the dispersion of contaminants [16]. The wind direction and speed and the vertical atmospheric stability, contingent on the air temperature, are considered the most important and decisive meteorological dispersion conditions. The wind characteristics influence the transport of pollutants in the horizontal direction, the atmospheric stability and the air temperature influence the dispersion of pollutants in the vertical direction. In the most stable situations, the air temperature rises with the height and the conditions for the vertical mixing are the worst. Conversely, at unstable stratification, the temperature drops more quickly than it would under regular conditions in the atmosphere, and the conditions for pollutant dispersion are favorable [17–19].

It is possible to express the dispersion conditions numerically using the ventilation index, which is defined as a product of the mixing layer height and the average wind speed in the mixing layer [20]. A situation with unfavorable dispersion conditions does not always have to mean the occurrence of high contaminant concentrations. It is important to bear in mind the situation duration, the original pollution level, the source distribution, and the emissions into the layer under inversion. It is, however, possible to state that a significant exceedance of the allowed pollution limits occurs mostly in mildly unfavorable and unfavorable dispersion conditions, or due to the concurrence of other meteorological factors. In order to calculate the ventilation index, the ALADIN (Aire Limitée, Adaptation Dynamique, Development International) numerical forecast model is used in the Czech Republic; this model is meant mostly for short-term forecast composition [21]. It is also possible to find out the information about the atmospheric boundary layer and the mixing layer height by means of direct monitoring. Measurements using radiosondes, high masts, and special devices, such as SODARs (Sonic Detection and Ranging), LIDARs (Light Detection and Ranging), and wind profilers, are used. In the presented analysis, mixing layer height measurement results obtained by means of ceilometers were used. The higher the mixing layer height, the more intensive the mixing of air, and thus, the better the conditions are for contaminant dispersion [22,23].
Awareness of the links between PM\textsubscript{10} concentrations and meteorological variables and their mutual combination is an important premise for assessment of the pollution situation and a possible pollutant concentration trend forecast [24–28]. The detection of these dependences is especially important for the assessment of complex situations with the occurrence of smog situations [29,30]. Apart from those pollutant concentrations exceeding the limit value, knowledge of the development of the meteorological situation, weather forecasts, and thus the possibilities for overall improvement or deterioration of the situation due to better or worse air pollutant dispersion, are important aspects for the announcement of smog situations. Improving the knowledge of correlations between the mostly high suspended particulate matter concentrations and the meteorological conditions may be a significant aspect for defining the measures preventing health hazards in the assessed densely populated region [31].

With this article, the authors reassume a similar topic of PM\textsubscript{10} concentration dependences on the mixing layer height in the area of interest [32]. In the mentioned analysis, the measurement results were assessed at the Věřňovice, Trinec-Kosmos, and Mošnov/Studénka sites for the cold parts of the years 2016/2017 and 2017/2018. The differences in the established facts were caused mostly by a different site location (orography) and by the structure of the sources that influence the pollution. Pollutants other than PM\textsubscript{10} and other meteorological variables were not included in the assessment.

The anticipated findings of this analysis shall be an understanding of the combinations of the meteorological conditions, the mixing layer height, and the connection to PM\textsubscript{10} concentrations. The highest PM\textsubscript{10} concentrations are generally reached at a low mixing layer height, low wind speed or calm air, and at a low air temperature. These assumptions correspond to previous results concerning the dependence of pollutant concentration on the mixing layer height in Europe [33–37]. Furthermore, the mixing layer height and its influence on the pollutant concentrations in the air have been written about in articles in Asia, e.g., in India [38] and China [39], which, as far as the comparability of pollution-meteorology correlations is concerned, is hardly transferrable, given these significantly different areas. The reasons are the different climatological conditions between the Czech-Polish border area and Asia, specifically the meteorological and geographical conditions of the selected locations, the air pollution intensity, and the different position and types of air pollution sources with respect to the monitored ambient air quality stations. The wind direction and speed, as well as the mixing layer height, also make it possible, at least in part, to reveal the main air pollution sources in the monitored border locations. With regard to a different area relief in which the monitored locations are situated, it has been assumed that the PM\textsubscript{10} concentration dependences on meteorological variables are bound to differ up to a certain point.

2. Experiments

For the assessment, sites of the Czech Hydrometeorological Institute [40] have been chosen—Trinec-Kosmos [41] and Věřňovice [42], located in the northeastern part of the Moravian-Silesian Region of the Czech Republic (Figure 1). The monitored sites are about 34 km apart in a straight line. Both sites belong to the National Air Quality Monitoring Network of the Czech Republic [43]. At the Trinec-Kosmos and Věřňovice sites, PM\textsubscript{10} concentrations are measured by means of a radiometric method based on the absorption of beta radiation in a sample collected on filter material by means of an Automatic suspended particulate monitor MP101M [44]. The wind direction and speed are measured at a standard height of 10 m by means of a WindSonic machine, based on ultrasound technology. The air temperature is scanned at a height of 2 m above the ground, by means of a Cormet company machine. At both sites, special ceilometer devices are placed that are traditionally used to measure the height of the cloud base and the cloud amount in individual layers, the vertical visibility, and the aerosol concentration in the ground atmospheric layer. These ceilometers contain a presentation module that enables the measurement and display of the structure of the atmospheric boundary layer based on an algorithm that determines the thickness of the mixing layer depending on the aerosol concentration in the atmosphere. In both cases, these are Vaisala CL31 Ceilometers [45].
It is necessary to point out that measurement by means of ceilometers is limited by height. Although the upper measurement limit reaches the height of a few kilometers, it is impossible to regard the Earth’s surface level as the bottom limit. In the vicinity of the Earth’s surface, the measurement is limited by higher noise that can reach up to a height of 50 m [46].

![Figure 1](image1.png)  ![Figure 1](image2.png)

**Figure 1.** The assessed site location: (a) The location within Central Europe; (b) The location within the Czech-Polish border region.

For the purposes of the analysis, the average hourly values of the mixing layer height were used, calculated from the data measured at 16-second intervals from the ceilometers at the Trinec-Kosmos and Věřňovice sites. The output hourly values of PM$_{10}$, temperature, wind speed, and direction were also used in the analysis. Due to large diurnal mixing layer height (MLH) variation, the analysis also included the daily averages of the parameters. Correlation coefficients in the analysis are calculated from the hourly and daily data in relation to the MLH, which was divided into 20 intervals and, subsequently, from their arithmetic means and medians. In the case of daily MLH averages, intervals from a height of 150 m are available. The assessed period is from 2016 to 2019. The data utilization rate used for the monitored period was high. In the case of hourly data on PM$_{10}$ and meteorological variables (air temperature, wind direction, and speed) the average at both sites is 98%; it is 87% from the ceilometers’ data for the mixing layer height. In order to assess higher PM$_{10}$ concentrations and exceedance numbers, a limit value of 50 µg·m$^{-3}$ was purposefully chosen, which corresponds to the value of the daily pollution limit for PM$_{10}$ [13]. The term “cold period of the year” is used for the months of January through March and October through December; the term “warm period of the year” is used for the months of April through September.

In the analysis, concentration roses were also used for illustrative purposes, which show average PM$_{10}$ concentrations for the given wind direction and speed, as well as weighted concentration roses [47]. The difference between a concentration rose and a weighted concentration rose is that the latter provides information about how often a given wind direction and speed combination occurs and states to what extent the concentrations detected for a given wind speed and direction affect the overall average concentration for a given period. The comparison of the two roses may show a significant pollution source located, however, in a sector from which the wind only rarely blows and thus does not contribute significantly to the overall average concentration. The concentration rose reveals what the pollution situation was at maximum concentrations of a given contaminant at a particular site; the weighted concentration rose shows from what wind direction and at what speed the pollution came to the largest extent for the whole period. For the above-described reasons, both rose types for
The same site and period may vary significantly [48]. In the presented roses, there is information about calm air, which is a situation with a wind speed of 0–0.2 m s\(^{-1}\).

The Třinec-Kosmos site is situated in the center of the city of Třinec, in a typical residential development. In the immediate vicinity of the site there are a parking lot as well as a residential development type of road. In the northeastern (NE) direction from the site, the closest pollution line source is a frequented road with an average number of vehicles of more than 9000 every 24 h [49]. The Třinecké železárny a.s. industrial plant premises are situated in the northwestern (NW) sector about 1.5 km from the site. The closest housing development with individual heating is situated in the N–NE direction about 500 m from the site, another is in the S–SW direction about 1 km from the site. The shortest route towards the Polish border is in the NE direction about 3 km from the site.

The Třinec-Kosmos site is situated at the end of the Jablunkov Furrow, which starts at Jablunkov Pass and separates the Moravian-Silesian and the Silesian Beskydy mountain ranges. The shortest distance to the foothills of the Beskydy mountains is just 4.5 km southwest of the station. In the furrow is the Olše River basin that drains off the surrounding undulating relief [50,51]. The direction of the furrow also determines the prevailing wind direction along the southeast/northwest axis (Figure 2a) [52].

In the vicinity of the site, at a distance of 40 m in the NE direction, there is a high-rise, a roughly 50-meter-high building that may, with regard to the prevailing wind direction, slightly muffle the wind influence from the northeast.

![Figure 2](image1)

**Figure 2.** Average frequency of counts by wind speed and wind direction, 2016–2019: (a) Třinec-Kosmos; (b) Věřňovice.

According to the existing pollution assessment analyses of the Třinec-Kosmos site, apart from the significant contribution of secondary particles, a combination of sources from abroad and local heating and industry, as well as road traffic, contribute significantly to the PM\(_{10}\) concentrations [52]. At lower wind speeds of up to 2 m s\(^{-1}\), with a prevailing southeastern (SE) wind direction, a significant influence of dense housing development with individual heating is very likely to be seen at the site. Conversely, in situations with a NW wind direction, a significant influence of the industrial sources of Třinecké železárny is assumed, which includes both high- as well as low-emitting sources. At higher wind speeds, the influence of sources situated north of the site at a greater distance is also apparent here in the cold period of the year, which points at foreign sources from Poland.

The Věřňovice site is situated in the cadastral area of the Dolní Lutyně municipality, about 300 m east of the Věřňovice village and about 1 km south of the Czech-Polish border. In the immediate vicinity of the site, there is an infrequently used and partly paved road flanked by deciduous trees and
fields. Roughly 1.5 km in the NW direction, there is a highway with roughly 11,000 vehicles a day [49]. The closest significant industrial source, Elektrárna Dětmarovice a.s., is situated about 3 km southeast of the site. The Věřňovice site is situated in open terrain in the vicinity of the River Olše. At the site, the prevailing wind direction is along the southwest/northeast axis where the more frequent flow is southwest (Figure 2b). Věřňovice is situated in the area of the Ostrava Basin where the “Moravian Gate” ends; this divides the Podbeskydská hilly area and Nízký Jeseník [50,51]. The shape of the gate along the southwest/northeast axis significantly determines the prevailing wind direction of the region where Věřňovice is situated. It is apparent from the existing analyses that local fireplaces have the greatest contribution to the exceedance of PM$_{10}$ pollution limits here among the primary sources, mostly those used on the territory of Poland [52].

3. Results

The suspended particulate matter concentrations at both sites are assessed depending on the mixing layer height and other meteorological variables—air temperature, wind speed, and direction.

3.1. Dependence of Suspended Particulate Matter Concentrations on the Mixing Layer Height

Based on the average daily courses of hourly concentrations of PM$_{10}$ and MLH in the period between 2016 and 2019 for both stations, it is obvious that the lowest MLH is measured in the evening and night hours and the early morning hours (Figure 3). During this period, the highest PM$_{10}$ concentrations were observed at the Věřňovice station. In both cases, higher PM$_{10}$ concentrations were observed in the cold part of the year, when the daily variability was also greater than in the warm part of the year.

![Figure 3](image_url)

**Figure 3.** Daily course of hourly PM$_{10}$ concentrations and mixing layer height (MLH), 2016–2019: (a) Třinec-Kosmos; (b) Věřňovice.

The distribution of the PM$_{10}$ and MLH hourly values is depicted in Figure 4.

In order to work out the dependence of PM$_{10}$ concentrations on the mixing layer height over the period from 2016 to 2019, correlation coefficients have been calculated for the values of their averages and medians. The median calculations have been added for the purposes of comparison without the influence of extreme values. The highest mean and median hourly PM$_{10}$ concentrations at the Třinec-Kosmos station for the entire period between 2016–2019 were observed at MLHs of 150 to 300 m. At the Věřňovice station, the highest hourly concentrations were observed at the lowest MLH, below 200 m. Domestic heating, a low (in terms of height) source of air pollution with the typical chimney height below 15 m, has a significant effect at both locations. The Třinec-Kosmos station is also significantly affected by the large industrial site of Třinecké železárny (Třinec Ironworks), with a chimney height of 100 m or more. At MLHs below 300 m, there is an increase in PM$_{10}$ concentrations, due to emissions from both domestic heating and industrial sources. At lower MLHs, the emissions from higher chimneys spread above this height. At higher MLHs (below 300 m), both domestic heating emissions and nearby industrial sources are very significant. After comparing all the levels of MLH,
higher PM$_{10}$ concentration values occurred at Věřňovice than in Třinec-Kosmos. On the occurrence of higher levels of the mixing layer, lower PM$_{10}$ concentrations were reached at both sites (Figure 5). For the Třinec-Kosmos station, the correlation coefficient for the 2016–2019 period calculated from the mean hourly values of concentrations at increasing MLHs is −0.83; from the median values, it is −0.84. The correlation coefficient from the mean daily values is −0.79; from the median values, it is −0.77. At the Věřňovice station, the correlation coefficient from the mean hourly values for PM$_{10}$ in relation to the increasing MLH is −0.76, and from the median values, it is −0.85. The correlation coefficient at the Věřňovice station for the entire period from 2016 to 2019 from the mean daily values is −0.74; for the median values, it is −0.75.

**Figure 4.** Hourly PM$_{10}$ concentration variability in relation to MLH, 2016–2019: (a) Třinec-Kosmos; (b) Věřňovice.

**Figure 5.** Mean and median PM$_{10}$ concentration values at mean and median MLH values, Třinec-Kosmos, Věřňovice, 2016–2019.

The correlation coefficients (hourly and daily values) of PM$_{10}$ dependences on MLH differ over the period from 2016 to 2019 at both sites; it is, however, possible to state that they are statistically significant dependences in all the cases (Table 1). When comparing the MLH arithmetic and median averages in the individual years, the highest values were reached in 2019 at both sites; they were higher at Věřňovice than at Třinec-Kosmos (Table 2).
3.2. Dependence of the Suspended Particulate Matter Concentrations on the Mixing Layer Height and the Air Temperature

The distribution of temperature and MLH hourly values are depicted in Figure 6.

![Figure 6](image-url)

Table 1. Correlation coefficients of PM$_{10}$ concentration dependences on MLH and the mixing layer height values at the Trinec-Kosmos and Veľňovice sites, 2016–2019.

| Year | Trinec-Kosmos | Veľňovice  |
|------|---------------|------------|
|      | Hour          | Day        | Hour          | Day        |
|      | Arithmetic Mean | Median    | Arithmetic Mean | Median    | Arithmetic Mean | Median    |
| 2016 | -0.85         | -0.83   | -0.81         | -0.78     | -0.76         | -0.84     | -0.70 | -0.66 |
| 2017 | -0.81         | -0.83   | -0.84         | -0.81     | -0.80         | -0.90     | -0.72 | -0.69 |
| 2018 | -0.81         | -0.85   | -0.79         | -0.81     | -0.82         | -0.87     | -0.81 | -0.79 |
| 2019 | -0.84         | -0.84   | -0.72         | -0.69     | -0.64         | -0.78     | -0.73 | -0.85 |

Table 2. The mixing layer height values at the Trinec-Kosmos and Veľňovice sites, 2016–2019.

| Year | Trinec-Kosmos | Veľňovice  |
|------|---------------|------------|
|      | Arithmetic Mean (m) | Median (m) | Arithmetic Mean (m) | Median (m) |
| 2016 | 604           | 369        | 555           | 327        |
| 2017 | 659           | 425        | 628           | 372        |
| 2018 | 675           | 450        | 670           | 408        |
| 2019 | 705           | 501        | 821           | 553        |

3.2. Dependence of the Suspended Particulate Matter Concentrations on the Mixing Layer Height and the Air Temperature

The distribution of temperature and MLH hourly values are depicted in Figure 6.

![Figure 6](image-url)

Figure 6. Variation of hourly temperature values in relation to MLH, 2016–2019: (a) Trinec-Kosmos; (b) Veľňovice.

The lowest average and median temperature values (hourly and daily values) were reached at both sites over the whole period from 2016 to 2019 on the occurrence of MLHs of 150 up to 300 m. From an MLH of 300 m the mean temperature values rise with increasing MLHs (or range around similar values) and the highest values are observed at an MLH above 800 m (Figure 7). On average, the statistical dependency expressed by a correlation coefficient between the temperature and MLH is significant over the whole period from 2016 to 2019 at both sites. The correlation coefficients calculated from the mean and median values (from the hourly and daily data) range between 0.85 and 0.94 on average for the entire period of analysis for both stations.
The correlation dependences between the temperature and the mixing layer height changed over the years of 2016 to 2019 at both sites. The correlation dependence is significant every year of the monitored period (Table 3).

Table 3. Correlation coefficients of dependences of the temperature on the MLH at the Třinec-Kosmos and Věřňovice sites, 2016–2019.

| Year | Třinec-Kosmos | | | Věřňovice | | |
|------|---------------|-----------------|-----------------|-----------------|-----------------|
|      | Hour          | Day             | Hour          | Day             | Hour          | Day             |
|      | Arithmetic Mean Median | Arithmetic Mean Median | Arithmetic Mean Median | Arithmetic Mean Median |
| 2016 | 0.95          | 0.88            | 0.89          | 0.81            | 0.94          | 0.87            |
| 2017 | 0.92          | 0.84            | 0.92          | 0.94            | 0.96          | 0.91            |
| 2018 | 0.90          | 0.82            | 0.88          | 0.90            | 0.93          | 0.84            |
| 2019 | 0.96          | 0.92            | 0.85          | 0.82            | 0.91          | 0.79            |

The PM$_{10}$ concentrations are significantly dependent on the mixing layer height and temperature in the case of both sites over the entire monitored period from 2016 to 2019. The average PM$_{10}$ concentrations reach their highest values at temperatures below 1 °C. At the Třinec-Kosmos location, the highest average daily and hourly PM$_{10}$ concentrations were observed at temperatures below 1 °C and a mixing layer height lower than 200 to 300 m. At temperatures above 5 °C, the average daily PM$_{10}$ concentrations were highest at an MLH between 200 and 300 m. At a higher MLH the PM$_{10}$ concentration values were almost identical. The average daily and hourly PM$_{10}$ concentrations at the Věřňovice station reached the highest values during the entire period of analysis in combination with the relationship between the temperature and height of the mixing layer, at temperatures below 1 °C and a mixing layer height below 200 m. At the same temperature, but at an MLH of 200–300 m, the average daily concentrations were lower, but still double the limit value of 50 µg·m$^{-3}$. A more significant PM$_{10}$ concentration dependence on temperature as well as MLH was observed at Věřňovice. At Třinec, a lower temperature had a substantial influence on PM$_{10}$ concentrations; the mixing layer height influence was shown mostly up to an MLH of 400 m. The combination of a low temperature and low MLH level corresponds to the general assumption about the occurrence of poorer dispersion conditions in the cold period of the year as opposed to the assumed occurrence of better dispersion conditions in the warm period of the year (Figures 8a and 9a). When observing the correlations among the numbers of exceedances of PM$_{10}$ daily concentration values of 50 µg·m$^{-3}$ and the MLH and temperature (Figures 8b and 9b), a decreasing count of exceedances of this value is apparent with
a rising temperature and increasing MLH. The highest number of 50 \( \mu g \cdot m^{-3} \) value exceedances for PM\(_{10} \) daily concentrations is observed at temperatures below 1 °C and MLHs below 300 m. For both sites, a slight increase in the number of 1-hour PM\(_{10} \) concentrations above 50 \( \mu g \cdot m^{-3} \) is apparent at an MLH level above 1000 m as opposed to 800 to 1000 m.

![Figure 8](image1.png)

**Figure 8.** The Trinec-Kosmos site, 2016–2019: (a) Average PM\(_{10} \) concentrations depending on the mixing layer height and temperature; (b) Exceedance count of daily concentrations of 50 \( \mu g \cdot m^{-3} \) depending on the mixing layer height and temperature.

![Figure 9](image2.png)

**Figure 9.** The Věřnovice site, 2016–2019: (a) Average PM\(_{10} \) concentrations depending on the mixing layer height and temperature; (b) Exceedance count of daily concentrations of 50 \( \mu g \cdot m^{-3} \) depending on the mixing layer height and temperature.

### 3.3. Dependence of Suspended Particulate Matter Concentrations on the Mixing Layer Height, Wind Speed, and Direction

The distribution of wind speed and MLH hourly values are depicted in Figure 10. At the Trinec-Kosmos site, a wind speed measured at 10 m does not significantly change with a rising MLH. At the Věřnovice site, the wind speed change is more apparent with a rising MLH. Here, the difference of the average values of the wind speed on the occurrence of MLH levels of up to 200 m and above 400 m makes up 1 and more m\( \cdot \)s\(^{-1} \) (Figure 11). The correlation dependences of the wind speed on MLH are low at the Trinec-Kosmos site; the Věřnovice site shows a statistically more significant dependence than the Trinec-Kosmos site. Higher wind speeds are observed at the Věřnovice station, compared to the Trinec-Kosmos station. The differences are due to the more complex orography of the Trinec region compared to the relatively flat and open region surrounding Věřnovice. The station in Trinec is located in a housing development, at the foothills of the Beskydy Mountains (Chapter 2). This means that the correlation between MLH and wind speed at the Trinec-Kosmos station is much less significant or even insignificant, compared to the station in Věřnovice.

The correlation dependences of the wind speed and MHL also differ from site to site throughout the years of the assessed period. A statistically more significant dependence of the wind speed on the MLH is apparent at the Věřnovice site (Table 4).
For PM10 daily concentrations is observed at temperatures below 1 °C... Speed, and Direction
The distribution of wind speed and MLH hourly values are depicted in Figure 10.

(a) 
(b)

Figure 10. Variation of hourly wind speed values in relation to MLH, 2016–2019: (a) Trinec-Kosmos; (b) Veřňovice.

Figure 11. Average and median values of the wind speed at average and median MLH values, Trinec-Kosmos, Veřňovice, 2016–2019.

Table 4. Correlation coefficients of dependences of the wind speed on the MLH, Trinec-Kosmos and Veřňovice, 2016–2019.

| Year | Trinec-Kosmos | Veřňovice |
|------|---------------|-----------|
|      | Hour          | Day       | Hour          | Day       |
|      | Arithmetic Mean | Median | Arithmetic Mean | Median | Arithmetic Mean | Median | Arithmetic Mean | Median |
| 2016 | 0.53          | 0.45     | 0.28          | 0.32     | 0.69          | 0.69   | 0.07           | 0.32   |
| 2017 | 0.56          | 0.41     | 0.46          | 0.55     | 0.79          | 0.79   | 0.53           | 0.49   |
| 2018 | 0.60          | 0.53     | 0.1           | 0.27     | 0.75          | 0.76   | 0.52           | 0.35   |
| 2019 | 0.51          | 0.34     | 0.01          | 0.2      | 0.63          | 0.71   | 0.43           | 0.37   |

PM10 concentrations are significantly dependent on the MLH and wind speed. At the Trinec-Kosmos site, situations with calm air and wind speeds of up to 0.5 m·s⁻¹ had an apparent influence on higher hourly PM10 concentrations. At wind speeds between 0.5–2 m·s⁻¹, the highest average hourly concentrations were observed at MLHs of 200–300 m. The average daily PM10 concentrations were always the highest at MLHs below 200 m. Average daily wind speeds below 0.5 m·s⁻¹ were not available for the two stations of interest (Trinec-Kosmos and Veřňovice). The daily average wind speed values were 0.5 m·s⁻¹ or more. The influence of the wind speed on PM10 concentrations decreased with the mixing layer height at the Trinec-Kosmos site. At wind speeds above 2 m·s⁻¹, the PM10 concentrations were lowest at the site. During the observation of the occurrence
of the hourly PM$_{10}$ concentrations above 50 µg·m$^{-3}$, the highest count of these higher values was undoubtedly at wind speeds of 1–2 m·s$^{-1}$ and 0.5–1 m·s$^{-1}$ and MLHs of up to 300 m. The lowest number of daily PM$_{10}$ concentrations above 50 µg·m$^{-3}$ was observed at wind speeds above 2 m·s$^{-1}$ and MLHs above 400 m (Figure 12).

![Figure 12](image1.png)

**Figure 12.** The Trinec-Kosmos site, 2016–2019: (a) Average PM$_{10}$ concentrations depending on the mixing layer height and wind speed; (b) The count of exceedances of daily concentrations of 50 µg·m$^{-3}$ depending on the mixing layer height and wind speed.

The situation at the Věřňovice site was slightly different. The highest average hourly PM$_{10}$ concentrations were observed at MLHs below 200 m and wind speeds between 0.5–2 m·s$^{-1}$, or under calm conditions and wind speeds below 0.5 m·s$^{-1}$. The highest average daily PM$_{10}$ concentrations were observed at MLHs below 200 m and wind speeds below 2 m·s$^{-1}$. The lowest average PM$_{10}$ concentrations were reached at all levels with the highest wind speeds above 2 m·s$^{-1}$. The highest number of 50 µg·m$^{-3}$ value exceedances for PM$_{10}$ daily concentrations was observed at wind speeds between 1 and 2 m·s$^{-1}$ and MLHs of 200 to 300 m, at lower wind speeds below 1 m·s$^{-1}$, and MLHs below 200 m (Figure 13).

![Figure 13](image2.png)

**Figure 13.** The Věřňovice site, 2016–2019: (a) Average PM$_{10}$ concentration depending on the mixing layer height and wind speed; (b) The count of exceedances of daily concentrations of 50 µg·m$^{-3}$ depending on the mixing layer height and wind speed.

The prevailing wind directions vary noticeably over the whole period from 2016 to 2019 at both sites (Figure 14). At the Trinec-Kosmos site, the prevailing wind direction is along the southeast/northwest axis, where the SE wind direction prevails. The average relative frequencies of counts of wind directions changed in the individual months throughout the whole period; the prevailing SE wind direction, however, remained. The least frequent wind direction was NE. At the Věřňovice site, the highest frequency of counts of the wind direction is along the southwest/northeast axis; the frequencies of counts of the prevailing wind direction, however, differed throughout the year. On average, the southwestern
(SW) wind direction prevailed in the cold months (October to March). In the warm part of the year, the NE and NW wind directions prevailed. The SE wind direction at the Věřňovice site was the least frequent throughout the whole monitored period.

![Image](a)

![Image](b)

**Figure 14.** Relative frequency of counts of wind direction representation, 2016–2019: (a) the Třinec-Kosmos site; (b) the Věřňovice site.

The average monthly PM$_{10}$ concentrations reached the highest values at the lowest average MLH in the cold months of the year at all wind directions at Třinec-Kosmos. The lowest average monthly PM$_{10}$ concentrations were reached during the SE wind direction. During the SE and E wind directions, the MLHs also reached the lowest values of the average monthly heights of the mixing layer. In comparing the years from 2016 to 2019, it is apparent that on average, 2019 showed higher values of the mixing layer height than the preceding years, which is also reflected in the lower average PM$_{10}$ concentrations (Figure 15).

![Image](c)

![Image](d)

![Image](e)

![Image](f)

**Figure 15.** Average PM$_{10}$ concentrations depending on the mixing layer height and wind direction, the Třinec-Kosmos site, 2016–2019.

The weighted concentration roses for PM$_{10}$ divided according to the mixing layer height (Figure 16) show that the most frequent contribution to the average PM$_{10}$ concentrations for the period from 2016 to 2019 comes to the Třinec-Kosmos site from a southeasterly direction, which basically corresponds to the described prevailing wind directions at this location. This information, however, does not necessarily mean that the maximum PM$_{10}$ concentration contributions also come from the same direction; winds from other directions can carry higher concentrations of PM$_{10}$, which may lead to the exceedance of the
daily limit of the pollution value for PM$_{10}$ of 50 µg·m$^{-3}$. The division of the weighted concentration roses according to the mixing layer height describes the PM$_{10}$ concentration at the height of the ground measurement at the site when the mixing layer height reached various levels.

Figure 16. PM$_{10}$ weighted concentration rose divided according to the mixing layer height (in m), the Třinec-Kosmos site, 2016–2019.

The lowest MLHs at the Věřňovice site are registered for the NE and E wind directions, when the highest average PM$_{10}$ concentrations were reached over the entire period from 2016 to 2019. Conversely, the highest MLHs are reached during wind directions from the S to W sectors, when the lowest average monthly PM$_{10}$ concentrations occur during the SE and SW wind directions. Having compared the individual years, it becomes apparent that as well as at Třinec-Kosmos, the average MLH in 2019 was, on the whole, the highest, and the average PM$_{10}$ concentrations reached the lowest values of the entire monitoring period (Figure 17).

Figure 17. Average PM$_{10}$ concentrations depending on the mixing layer height and wind direction, the Věřňovice site, 2016–2019.
The weighted concentration roses for PM$_{10}$, divided according to the mixing layer height (Figure 18), show that the most frequent contribution to the average PM$_{10}$ concentrations for the period from 2016 to 2019 come to the Věřňovice site from the NE sector at wind speeds of up to 2 m·s$^{-1}$, despite the fact that, on the whole, the SW wind direction significantly prevails at the site. The division of the weighted concentration roses according to the mixing layer height does not describe the PM$_{10}$ concentrations and the wind direction at a certain MLH, but a situation at a ground measurement height at the site and during the time when the mixing layer height reached the described height.

![Figure 18](image)

**Figure 18.** A weighted PM$_{10}$ concentration rose divided according to the mixing layer height, the Věřňovice site, 2016–2019.

When comparing Figures 14, 17 and 18, a difference in the layout of the frequency of counts of wind directions between both sites becomes apparent. While at the Třinec-Kosmos site, a boundary of the prevailing wind direction along the southeast/northwest axis is more sharply visible in the pictures, at Věřňovice, despite the fact that the prevailing wind direction is along the southwest/northeast axis, the frequencies of counts of directions are spread more evenly to the other sectors, as well. This situation corresponds to the information described in Chapters 1 and 2.

The concentration roses (Figure 19) for 2016–2019 show that the maximum PM$_{10}$ concentration contributions come to the Třinec-Kosmos site at a low MLH level of up to 200 m and low wind speeds of up to 1 m·s$^{-1}$ from the northern (N) to the western (W) sector. For MLHs of 200–400 m, an influence of the wind direction from the E sector is also significantly shown, as well as the influence from the N and W sectors. The average maximum PM$_{10}$ concentration contributions at the site occur at MLHs of up to 400 m above the ground and at wind speeds of up to 2 m·s$^{-1}$. On the occurrence of a higher MLH level of up to 600 m, a higher wind speed of above 4 m·s$^{-1}$ from the N and NE directions contributes to the resulting situation more significantly. From the point of view of the PM$_{10}$ concentration roses divided according to the mixing layer height for the Třinec-Kosmos site, it is possible to regard the SW wind directions as those directions with the lowest PM$_{10}$ contribution; at higher wind speeds of above 4 m·s$^{-1}$, the S-SE directions also contribute the lowest amount.

Figure 20 illustrates concentrations for the Věřňovice site for 2016–2019, which are divided according to the different heights of the mixing layer. The highest PM$_{10}$ concentration contributions are reached up to the occurrence of MLHs of 200 m above the ground at wind speeds of up to 2 m·s$^{-1}$ from the E–NE. The influence of the wind from these directions is shown in the average maximum PM$_{10}$ contributions up to the height of 600 m, during which time the influence of a rising wind speed of up to about 6 m·s$^{-1}$ also increases proportionally.
When comparing the PM$_{10}$ concentration contributions at the sites after the division according to the wind directions and according to the MLH levels (Figures 15–20), it is apparent that at higher MLHs the contributions generally reach lower values for all wind directions.

4. Discussion

The detection of dependences of PM$_{10}$ concentrations on the mixing layer height may point out at possible pollution sources in the area; it would, however, be necessary to check the exact determination of the source type by other methods used in assessing the field of air quality, e.g., by more detailed air pollutant measurements and by methods for pollution source identification, e.g., Positive Matrix Factorization (PMF) [53] and back trajectories [54]. The high PM$_{10}$ concentrations in combination with the low mixing layer height provide information about the high influence of low-emitting sources of
air pollution (individual heating) on the air quality at a given site. On the other hand, higher PM$_{10}$ concentrations reached at higher MLH levels show the influence of industrial sources, sources with emission release at greater heights. High concentrations combined with a low wind speed, perhaps even with no wind, show the influence of sources in the vicinity of the site. Conversely, higher wind speeds combined with high PM$_{10}$ concentrations correspond to the premise of transport from more remote places to the site [16,17,24–26].

The findings of this work correspond to the conclusions of Air Quality Improvement Program [52] elaborated for both sites and areas of interest in which the sites are located and also to the “Air Pollution Sources Contribution to PM$_{2.5}$ Concentration in the Northeastern Part of the Czech Republic” analysis [55].

As mentioned before in Section 2, the meteorological variables data used the analysis from the ground measurements of temperature at a height of 2 m above the ground; the wind direction and speed were measured at 10 m above the ground. The values of these variables at various mixing layer heights were not included in the analysis. A vertical wind direction and speed and temperature profile would be a significant addition to the assessment that would, among other things, provide information about the occurrence of temperature inversions. This information could be completed with distant measurement data or from model calculations in a subsequent assessment. It is necessary to point out again that the ceilometer measurement is biased by noise at the lowest level above the ground [46], which interferes up to roughly 50 m according to the existing available sources.

Although the PM$_{10}$ concentrations are significantly dependent both on the wind speed and the mixing layer height, the dependence of the wind speed and MLH does not show a significant statistical dependence. The differences are also apparent between both assessed sites with a statistically higher dependence detected at the Věřňovice site. Orography is assumed to play an important role in this aspect (Section 2). While the Věřňovice site is situated in open terrain, the Třinec-Kosmos site is situated in an urban development at the end of Jablunkov Furrow at the foothills of the Moravian-Silesian Beskydy mountains.

Correlation dependences for the wind direction and MLH have not been elaborated. The authors do not consider this aspect important. The prevailing wind direction is different at both sites, the difference lying mostly in the area geomorphology where the sites are situated. The authors find the wind direction and speed assessment sufficiently useful, divided according to the MLHs mentioned in this article. These indicators serve primarily for detection of the origin and location of possible pollution sources that influence the given area.

Another opportunity for analysis is the assessment of other pollutants’ dependence on the mixing layer height. Among the examples are ground-level ozone and processing for the warm part of the year [56], as well as a more detailed assessment of the behavior of pollutants such as sulfur dioxide, nitrogen dioxide, and suspended particles during winter smog situations when there is a low mixing layer height [32].

The disadvantage of an assessment of the relationships between meteorology and ambient air quality with regard to the mixing layer height lies in the fact that there are no available data from a longer sequence of ceilometer measurement. The assessment of the mixing layer height trends from a long-term point of view would surely bring additional information about the development or changes of the dispersion conditions, especially in the areas with poor air quality. The observed information may serve as a basis for the evaluation of current situations with high PM$_{10}$ concentrations, for announcing and invalidating smog situations, and model solutions of pollution situation forecasts.

5. Conclusions

The presented analysis provides evidence of the influence of the mixing layer height, combined with other meteorological variables, on PM$_{10}$ concentrations in the area of the Czech-Polish border, which faces problems with a complex representation of a substantial number of various pollution
source types. The influence on the pollution situation in combination with meteorological variables at a given site also differs depending on the varied geographical conditions.

The influence of a temperature below 1 °C on the pollution situation is noticeable at both sites. The highest numbers of exceedances of PM$_{10}$ daily concentrations of 50 µg·m$^{-3}$ occur at low temperatures combined with a mixing layer height of up to 300 m. In the case of Trinec-Kosmos, the average PM$_{10}$ concentrations at temperatures below 1 °C reach the highest values on the occurrence of a mixing layer height of up to 300 to 400 m. The decrease in PM$_{10}$ concentrations with increasing MLHs at temperatures below 1 °C is not very significant. In this case, the influence of a rising height of the mixing layer at temperatures below 1 °C on the average PM$_{10}$ concentrations is not as significant as in the case of Věřňovice, where a difference of several tens of µg·m$^{-3}$ in the average PM$_{10}$ concentrations is observed between levels of up to 200 m, and levels of 200–300 m.

The influence of the wind speed on the occurrence of various MLHs on PM$_{10}$ concentrations is also noticeable. The average PM$_{10}$ hourly concentrations at Trinec-Kosmos are the highest at wind speeds of up to 0.2 m·s$^{-1}$, or 0.5 m·s$^{-1}$, at MLH levels of up to almost 600 m, at Věřňovice the influence of wind speeds of up to 2 m·s$^{-1}$ is detected. In both cases, throughout the assessment of the number of PM$_{10}$ daily concentrations above 50 µg·m$^{-3}$, it is then clear that higher values occur in the case of Trinec at wind speeds of 0.5 up to 2 m·s$^{-1}$ and MLHs of 200–300 m. At the Věřňovice station, the highest number of limit value exceedances was observed at wind speeds between 1–2 m·s$^{-1}$ and at wind speeds above 2 m·s$^{-1}$ and MLHs of 200–300 m.

The influence of the wind direction combined with the MLH is different at both sites; however, it is crucial as well. Despite the fact that the most frequent PM$_{10}$ contributions come to the Trinec-Kosmos site from the SE direction, the average maximum concentration contributions come from the W–N sectors at low wind speeds and MLHs of up to 400 m. In Věřňovice, regardless of the prevailing SW wind direction, sources in the NE–E sector from the site have a crucial influence on the air pollution level caused by PM$_{10}$. The maximum PM$_{10}$ contributions are reached on the occurrence of MLHs of up to 200 m at Věřňovice station.

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