Critical Review: Propensity of Premise Plumbing Pipe Materials to Enhance or Diminish Growth of Legionella and Other Opportunistic Pathogens

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Abstract: Growth of Legionella pneumophila and other opportunistic pathogens (OPs) in drinking water premise plumbing poses an increasing public health concern. Premise plumbing is constructed of a variety of materials, creating complex environments that vary chemically, microbiologically, spatially, and temporally in a manner likely to influence survival and growth of OPs. Here we systematically review the literature to critically examine the varied effects of common metallic (copper, iron) and plastic (PVC, cross-linked polyethylene (PEX)) pipe materials on factors influencing OP growth in drinking water, including nutrient availability, disinfectant levels, and the composition of the broader microbiome. Plastic pipes can leach organic carbon, but demonstrate a lower disinfectant demand and fewer water chemistry interactions. Iron pipes may provide OPs with nutrients directly or indirectly, exhibiting a high disinfectant demand and potential to form scales with high surface areas suitable for biofilm colonization. While copper pipes are known for their antimicrobial properties, evidence of their efficacy for OP control is inconsistent. Under some circumstances, copper’s interactions with premise plumbing water chemistry and resident microbes can encourage growth of OPs. Plumbing design, configuration, and operation can be manipulated to control such interactions and health outcomes. Influences of pipe materials on OP physiology should also be considered, including the possibility of influencing virulence and antibiotic resistance. In conclusion, all known pipe materials have a potential to either stimulate or inhibit OP growth, depending on the circumstances. This review delineates some of these circumstances and informs future research and guidance towards effective deployment of pipe materials for control of OPs.

Keywords: non-tuberculous mycobacteria; Pseudomonas; Acinetobacter; amoebae; copper; iron; PEX; PVC; drinking water; disinfection

1. Introduction

Legionnaires’ Disease is the “leading cause of reportable waterborne illness” in the United States [1,2], with 52,000–70,000 cases per year [1,3,4], 8000–18,000 hospitalizations [5], an overall mortality rate of 15% [4], and high healthcare and legal costs [2,6–8]. Bacteria belonging to the genus Legionella are the causative agent of Legionnaires’ disease and Pontiac Fever, which infect the human respiratory system via inhalation or aspiration. Legionella is classified as “opportunistic”
because it preferentially infects those with underlying illnesses or weakened immune systems [4,8,9]. To date more than 60 Legionella species have been identified [10], with Legionella pneumophila being the species most commonly attributed to human disease [11]. Legionella can be found even in “the most aggressively treated drinking water” [12]. Studies have confirmed that potable water is a key source of infection [1,4,13–17], for both hospital- and community-acquired cases [18–20]. Other opportunistic pathogens (OPs) such as nontuberculous mycobacteria (NTM), Pseudomonas aeruginosa, and Acanthamoebae, can similarly be transmitted via tap water and tend to infect individuals belonging to certain risk groups [8].

To infect humans, Legionella and other OPs must be present in tap water at the point of use. While Legionella can occasionally survive drinking water treatment and be transported through the main water distribution system, the primary environment for Legionella proliferation to numbers needed to infect humans generally occurs in building or “premise” plumbing [21,22]. Premise plumbing includes the service pipe that connects buildings to the water main, in addition to the full array of components comprising cold and hot portions of a building’s potable water system [8]. Premise plumbing is characterized by high surface area to volume ratios, longer stagnation times, low disinfectant residual, areas with excess sediment and scale, chemically and biologically reactive plumbing materials, and water with relatively warm temperatures. Such conditions can create ideal micro- and macro-environmental niches for growth of various OPs [1,8,23].

Premise plumbing is a key conduit for human exposure via showering, handwashing, and other applications that create airborne aerosols [24]. Legionella has been detected in faucets, showerheads, decorative fountains, grocery store mist systems, ice machines, and cooling towers [13,14,16,25]. Larger buildings with more complex plumbing systems are more likely to create physicochemical conditions suited for Legionella proliferation, but it is also often detectable in water mains and residences with simple conventional hot and cold water plumbing systems [17,26,27]. A Centre for Disease Control (CDC) summary of Legionnaires’ Disease potable water outbreak investigations from 2000–2014, concluded that 85% of the cases had “deficiencies” in water system maintenance within buildings as a contributing factor [28] and that water chemistry flowing into buildings is one, but not the only, predictor of Legionella incidence [29,30].

The mechanisms by which premise plumbing influences L. pneumophila and other OPs, as well as the broader premise plumbing microbiome, are varied and complex (Figure 1). The influent water chemistry has been found to influence Legionella, and also strongly shape the plumbing microbiome, especially through the delivery of growth-promoting nutrients, growth-inhibiting disinfectants, and influent microorganisms [31–34]. The ecological interactions among microorganisms in biofilms of building plumbing systems can also help overcome barriers to growth from low nutrient levels and disinfectants [24,35,36]. Conversely, other interactions, such as competition, exclusion, predation, or inactivation of symbiotic organisms, may inhibit the growth of OPs [37]. The selective pressures in premise plumbing might also alter the physiologies of resident microbes in a manner that influences infectivity [38]. All these phenomena are further complicated by the fact that premise plumbing configurations, hydraulics, temperature, and water use patterns including velocity, flow or stagnation events, all differ significantly from building to building. In particular, there is strong variability due to occupancy, building size, water heater design, water saving devices, storage and other factors [39,40]. Thus, while there are many overarching similarities, every premise plumbing system is at least as variable as the occupants’ unique water use patterns and habits.
Figure 1. Overview of exemplar mechanisms by which pipe materials can affect OPs in premise plumbing. Depending on the circumstances, the pipe material itself can have direct effects on OPs growth by: (A) providing organic or inorganic nutrients that enhance growth, (B) acting as a growth-inhibiting antimicrobial, or (C) inducing viable-but-non-culturable (VBNC) status, from which microbes might recover in terms of infectivity and growth rates subsequent to exposure. Pipes can also indirectly affect OPs by: (D) consuming secondary disinfectants, allowing for microbial growth downstream, (E) evolving...
hydrogen gas or enhance nitrification, fueling autotrophic growth, or (F) developing thick pipe scales, which provide additional surface area for microbial growth, or (G) selecting for certain types of amoebae that are preferred hosts for bacterial OPs and protect them from negative effects of copper and disinfectants. Finally, pipes may unfavorably alter the physiology of microbes by increasing (H) OP virulence by selecting for resistance to phago-somal copper overload, or (I) resistance to antibiotics.

The type of pipe material can also strongly influence the relationship between premise plumbing materials and OPs through both direct effects (interaction with chemical species released from pipe) and indirect effects (secondary consequences of released material from pipes) by altering the level of nutrients, disinfectants, and microbial biomass (Table 1, Figure 1). Selection of pipe material can therefore strongly affect chemistry, biological stability [41], and microbiome composition [42] of the drinking water.

| Water Chemistry Attribute Influenced by Pipe Materials | Relevance to OPs | Effect of Pipe Materials on OPs Control as Mediated by the Indicated Water Chemistry Attribute |
|--------------------------------------------------------|------------------|-------------------------------------------------------------------------------------|
| Chlorine Disinfectant                                  | -                | [43]                                                                              |
| Chloramine Disinfectant                                | -                | [43,54]                                                                            |
| Assimilable Organic Carbon Carbon source               | 0                | [42,56,57]                                                                        |
| Hydrogen Gas (aq) Food web                             | 0                | [42,56,58,59]                                                                     |
| Release of Metals Release of metals                    | 0                | [59,62–64]                                                                        |

**Abbreviations:** OPs, opportunistic pathogens; PVC, Polyvinyl chloride; PEX, cross-linked polyethylene; SS, stainless steel; aq, aqueous. ¹Includes unlined iron and old galvanized iron pipes.

Motivations for this review include:

- Growing direct or indirect potable water reuse, which can sometimes alter levels of nutrients and Cu²⁺ in the source water [67].
- Increased natural organic matter (NOM) in some source waters as an indirect consequence of improving sulfur and nitrogen air pollution controls under rules and regulations such as the U.S. Clean Air Act or Directive 2008/50/EU [68–70].
- Emphasis on and investment in green building design for water and energy efficiency and associated unintended consequences for in-building hydraulics (e.g., more stagnation, higher surface area to volume ratios of water to plumbing surfaces, required hot water recirculation systems) that alter water chemistry and delivery of nutrients or disinfectants [39,54,71,72].
- Greater use of plastic pipes (e.g., PEX, PVC, polyethylene), which vary in leaching potential by type of plastic and due to the presence of proprietary stabilizers and processes [73].
- Increasing awareness of viable-but-non-culturable (VBNC) bacteria, which are difficult to measure directly. Molecular and fluorescence-based techniques suggest that they can be prevalent under certain circumstances [8,74] and recent evidence indicates they can still cause disease [75,76].
- Heightened concern about an array of bacterial OPs besides *Legionella*, including *Pseudomonas aeruginosa*, *Acinetobacter baumannii*, and NTM, as well as amoebae (e.g. *Acanthamoeba*, *Vermamoeba*), which can themselves be pathogenic or can serve as host organisms for bacterial OP proliferation [8].

Here we critically examine existing knowledge with respect to the direct (Section 2) and indirect (Section 3) effects of common metallic (copper, iron, zinc, aluminum, magnesium) and plastic (PVC, PEX) building pipe materials on the growth of *Legionella* and other OPs, in addition to identifying the complex effects of plumbing system configuration (Section 4) and the characteristics of the drinking water.
microbiome (Section 5). This review is particularly timely, at a moment when societal expectations for public health protection are elevated and expanding aspirations for improved water/energy conservation will be a major drive of water system design and pipe material selection [39]. In executing this review, we aimed to holistically assess the effects of pipe materials, primarily focusing on Legionella while including other OPs, seeking to shed light on why various pipe materials appear to sometimes enhance and other times diminish OP proliferation under real-world premise plumbing conditions.

2. Direct Effects of Plumbing Material on Pathogen Growth

2.1. Copper Has Both Antimicrobial and Micronutrient Properties

Copper is sometimes present at trace levels in the source water or in distributed water mains, but the main sources in premise plumbing are copper pipes and brass fittings that are installed beginning at the service line connecting the building to the water main (Figure 2). Due to long-lasting life span, durability, and relatively few concerns about metal release when compared to those of antiquated lead and galvanized iron alternatives, copper and its alloys are common in premise plumbing systems [77]. Copper is a registered antimicrobial of the US Environmental Protection Agency (EPA) [78] and listed as a biocidal product in the European Union, but some countries require special approval for use of copper in drinking water for OP control [79]. It is also an essential nutrient for all living organisms, including humans and OPs [59,80]. Here we review the mechanisms by which copper plumbing may influence control of various OPs (Table 2).

![Figure 2. Copper sources in premise plumbing [81–84]. Note that Cu-Ag Ionization systems can be used in either point of entry or hot water distribution networks.](image-url)
Table 2. Copper can be growth-promoting or -inhibiting to opportunistic pathogens.

| Opportunistic Pathogen | Associated Diseases | Exposure Route(s) | Inactivation via Copper | Growth via Copper |
|------------------------|---------------------|-------------------|------------------------|------------------|
|                        |                     |                   | Antimicrobial Efficacy * | Evidence for Cu-Induced VBNC | Micronutrient Activity | Amoeba-Mediated Growth |
| **Amoeba**             |                     |                   |                        |                  |                   |                       |
| **Acinetobacter baumannii** | Bacteremia, Meningitis, Pneumonia, Urinary tract infections [89] | Dermal, Inhalation [89] | Moderate to Somewhat inhibited [58,90–92] | Unknown | Yes | [93,84] |
| **Staphylococcus aureus** | Bacteremia, Endocarditis, Osteomyelitis, Pneumonia, Sepsis, Skin infections [95] | Dermal, Inhalation [96,97] | Moderate [90,98,99] | Unknown | Yes | [100,101] |
| **Stenotrophomonas maltophilia** | Bacteremia, Endocarditis, Eye infections, Meningitis, Pneumonia, Sepsis, Skin infections, Urinary tract infections [102,103] | Dermal, Inhalation [102,103] | Moderate [91,92] | Limited [104] | Yes | [105] |
| **Nontuberculous Mycobacteria (NTM): Mycobacterium avium complex; Mycobacterium abscessus complex; Mycobacterium kansasii and other species** | Bacteremia, Pneumonia, Skin infections [106] | Dermal, Ingestion, Inhalation [107] | Moderate [108–111] | Limited [109] | Yes | [36,112] |
| **Aeromonas hydrophila** | Gastroenteritis, Meningitis, Peritonitis, Pneumonia, Skin infections [113] | Ingestion, Inhalation [113] | Unknown [114] | Unknown | Yes | [115,116] |
| **Legionella pneumophila** | Legionnaires’ disease, Pontiac fever [117] | Inhalation [118] | Somewhat inhibited to High [62,83,119,120] | Moderate [121,122] | Yes | [125,124] |
| **Pseudomonas aeruginosa** | Bacteremia, Endocarditis, Eye infections, Gastroenteritis, Osteomyelitis, Pneumonia, Sepsis, Skin infections, Urinary tract infections [125] | Dermal, Ingestion, Inhalation [125,126] | Somewhat inhibited to High [90–92,96,99,127–132] | Strong [127,132,133] | Yes | [36,134] |

* Categorizations of efficacy based upon studies that showed planktonic phase growth inhibition at: <0.1 mg/L (High), 0.1–0.8 mg/L (Moderate), and >0.8 mg/L (somewhat inhibited) copper concentrations in water or media.
2.2. Copper Pipe as an Antimicrobial Material in Premise Plumbing

The antimicrobial properties of copper were first described more than 3000 years ago in the Hindu Vedas and are occasionally observed at least temporarily in modern plumbing systems [1,120,135–137]. The role of supplemental dosing of copper as disinfectants in building plumbing can be important, because Legionella and other premise-plumbing-associated OPs are more resistant to chlorine than traditional fecal-associated bacteria that are used for traditional water quality monitoring [8,24,138]. While there is no clear consensus on the primary mechanisms by which copper inactivates bacteria, two hypotheses have been put forward: (1) positively charged Cu\(^{2+}\) ions interfere with negatively charged cell membranes, creating holes; and (2) Cu\(^{2+}\) disrupts the replication and production of DNA, RNA, and proteins, potentially through metabolic cycling between Cu\(^{3+}\) and Cu\(^{2+}\) oxidation states, which generates radical oxidative species such as hydroxide radicals [139]. In potable water, copper passively released from plumbing materials can be present in the germicidal range for Legionella of 0.1–0.8 mg/L [62,119,120,140], even in some parts of plastic pipe systems connected with brass fittings [141,142]. Passive release or purposeful dosing that results in copper concentrations of 0.05–0.8 mg/L are thought to limit Legionella growth [62,83,119,120,143].

A number of studies have confirmed the efficacy of copper, either passively leached from premise plumbing materials [59,140,144] or actively added using copper-silver ionization (CSI) systems [62,83,145], as a Legionella antimicrobial. Biofilms grown at room temperature for 30 days in pre-sterilized reactors with copper, PVC, and stainless steel coupons were found to have lower total bacterial counts on copper than PVC surfaces [146]. Other batch reactor studies indicate similar results, demonstrating lower L. pneumophila numbers on copper plumbing than plastic plumbing [59,140,144,147]. Analogous responses to copper surfaces by other Ops, such as Klebsiella spp. [148], NTM [111,149], P. aeruginosa [128], and Aeromonas hydrophila [114], have been reported. Two different field studies found that copper concentrations were significantly lower in samples positive for L. pneumophila than samples negative for L. pneumophila [150,151]. Borella et al. [23,152] identified a threshold total copper level of 0.5 mg/L in one sample of water, above which samples were approximately two to seven times less likely to be positive for L. pneumophila.

Studies of CSI applications also demonstrate that copper can have direct antimicrobial effects. Lin et al. [83,109] showed that 0.5 and 48 h of exposure to 0.4/0.04 mg/L copper/silver achieved 99% inactivation of L. pneumophila and Mycobacterium avium, respectively, in bench-scale testing. Stout et al. [119] performed long-term monitoring of CSI systems in 16 hospitals and demonstrated their efficacy for Legionella control, as the numbers of hospitals with >30% Legionella positive samples dropped from 7/16 to 0/16, and no Legionnaire’s disease cases were reported in 15 out of 16 hospitals after the implementation of CSI. Addition of copper ions to solution from pipes or via CSI, at the bench and building-scale, has also been shown to inhibit the growth or reduce the frequency of OPs such as Staphylococcus spp.[98,99], Stenotrophomonas maltophilia [91,92,104], Acinetobacter baumannii [58,91,92], NTM [108,109], and P. aeruginosa [91,92,98,99,127,130].

2.2.1. Noteworthy Limitations to Copper’s Antimicrobial Efficacy

Despite the encouraging examples presented in the previous section, the overall success of copper as a disinfectant for Legionella is mixed [110]. Several studies have found that the antimicrobial effects of copper were limited, or that copper even encouraged growth of Legionella in some instances [63,83,122,153]. In one study, Legionella was consistently detected in a hospital hot water plumbing system with average pH = 7.7, even when copper was present at concentrations of 1.1 ± 0.2 mg/L [153]. Other studies have shown similar trends. For instance, Giao et al. [121] found no significant difference between biofilm formed on plastic (PEX and PVC) coupons and biofilms formed on copper coupons when the biofilms contained a heterogeneous community or when the biofilms were purely L. pneumophila. P. aeruginosa has been found to persist in hospital copper plumbing [129] and the implementation of a CSI system in one hospital did not appear to fully eliminate patient P. aeruginosa infections associated with exposures from faucets [130].
Prominently, in one field study conducted in Germany with low or no chlorine residual, hot water systems containing copper pipes were colonized with *Legionella* much more often (>30x) than those with galvanized steel or plastic pipes, despite the fact that the temperature of the hot water in these systems was similar. Also, samples (*n = 44*) from hot water recirculation lines with >0.5 mg/L of copper displayed 2,4000 ± 15,000 (mean ± standard deviation) CFU *Legionella*//L, while samples (*n = 153*) with <0.5 mg/L of copper had 10 ± 100 CFU *Legionella*/L [63].

There are many possible explanations for the apparent contradictions in overall impacts of copper (Table 2). It is important to first recognize that the antimicrobial properties of copper can be almost completely controlled by water chemistry (Figure 3). Notably, the concentration of Cu$^{2+}$ and its associated inorganic ions tend to decrease in concentration in aged pipes, at higher pH, or in the presence of common corrosion inhibitors, such as orthophosphate. Unfortunately, studies frequently do not collect or report such relevant data [63,129,130,153], limiting the ability to trace differences in copper’s antimicrobial efficacy to water quality parameters. There is also the likelihood of strain-to-strain differences in copper resistance, and the selection for copper resistant organisms in systems with copper pipes [154,155].

![Figure 3. Copper pipe corrosion and speciation is controlled by influent water chemistry and pipe age. Water chemistry parameters, such as pH, dissolved oxygen (DO), disinfectants, inorganic complexing agents (e.g., alkalinity, phosphate, and ammonia), organic complexing agents (e.g., natural organic matter (NOM)), hardness, trivalent metal ions (e.g., aluminum, iron), sulfate, and chloride can influence copper pipe dissolution, speciation, and the precipitation process. Copper is categorized as either free copper ions and inorganic complexed copper (considered relatively bioavailable), or organically complexed or particulate copper (considered relatively non-bioavailable). The level of copper species in the premise plumbing systems are also affected by the pipe use pattern (new vs. old pipes) and the water use pattern, including flow rate, stagnation and temperature.](image)

2.2.2. Water Chemistry Effects on Copper Bioavailability

The chemistry of the influent bulk water can reduce toxicity of copper by: (1) reducing overall solubility and the equilibrium level of Cu$^{2+}$ in the presence of copper rusts [156,157]; (2) forming copper complexes [158–160], (3) having elevated divalent (Ca$^{2+}$, Mg$^{2+}$) or trivalent (Fe$^{3+}$, Al$^{3+}$) cations, which compete with copper for uptake sites of organisms [161–163]. Therefore, water chemistry details are useful to explain the discrepancy of copper effects, but such information is often lacking in some studies [63,121,129,130,153].

Prior culture-based research demonstrated that precipitation of copper at pH 9 reduced toxicity of copper towards nascent *L. pneumophila* colonies by 16-fold relative to pH 7, where copper is more soluble [83]. Other compounds known to reduce levels of Cu$^{2+}$ by complexation and precipitation are logically expected to interfere with copper antimicrobial properties and include NOM and either ortho- or poly-phosphates [156–160]. Specifically, NOM and polyphosphate sequestrants can vary in concentration and complexation ability from water to water, can bind Cu$^{2+}$ and dramatically reduce its
bioavailability. Orthophosphate added as a corrosion inhibitor can reduce metal pipe corrosion rates and lower free metal ion concentrations in drinking water. For example, our research has shown that the addition of 3 mg/L of phosphate and 5 mg/L NOM at pH = 7 reduced copper’s antimicrobial effects towards *L. pneumophila* by four and seven times, respectively [164].

Copper’s antimicrobial properties are expected to increase at lower pH, lower hardness, lower Al<sup>3+</sup> and Fe<sup>3+</sup>, lower phosphate or polyphosphate, lower NOM, and colder temperatures due to known interactions with Cu<sup>2+</sup> ion. Studies of copper toxicity to algae and higher aquatic organisms have shown that Mg<sup>2+</sup>, Ca<sup>2+</sup>, Al<sup>3+</sup> and Fe<sup>3+</sup> compete with copper for binding sites, reducing the toxicity of copper [161–163]. For instance, Ebrahimpour et al. [161] reported that the 96-h median lethal concentration (LC50) values for *Capoeta fusca* increased roughly linearly (1.1 to 7.5 mg/L copper) over a hardness range of 40-380 mg/L as CaCO<sub>3</sub>. Trivalent metal ions, such as Al<sup>3+</sup> and Fe<sup>3+</sup>, can also form a layer of metal hydroxide gel around cells that can sorb copper and reduce its availability [165]. Free copper also tends to decrease at higher temperature and as pipe scales age [54,166].

2.2.3. Copper as a Nutrient in Premise Plumbing

Copper (Cu) is an essential micronutrient used in protein synthesis, respiration, various oxidation/reduction reactions and other functions in prokaryotes [80,167]. Accordingly, it is reasonable to suspect that copper piping might sometimes act as a source of this essential nutrient in premise plumbing, thereby increasing microbial growth relative to other materials. Buse et al. [122] showed that effluent from CDC biofilm reactors equipped with coupons of different pipe materials at pH > 8 and PO<sub>4</sub> > 0.2 mg/L, had up to 20× more *L. pneumophila* gene copies when copper coupons were used relative to PVC coupons. Mullis et al. [111] indicated that copper surfaces supported two to four times more *Mycobacterium abscessus* than PVC. Mathys et al. [63] reported that hot water systems containing copper pipes were colonized significantly more often than those with galvanized steel or plastic pipes.

2.3. Direct Release of Organic Carbon by Plastics

Potable water is oligotrophic, because organic carbon is relatively scarce and often limiting to the growth of drinking water microorganisms [24,168,169]. Plastic premise plumbing pipes, which are made with polymeric organic compounds, including stabilizers, flexibilizers and plasticizers, can leach organic carbon to water [56,57,170] whereas metallic pipes do not. These organic carbon compounds can fuel the growth of *Legionella* [45,59] and presumably other OPs. In some cases, the organics leached to water are not the polymers themselves, but rather are additives (i.e., flexibilizers, plasticizers, stabilizers) to improve aspects of pipe performance [42,170,171].

New PEX pipes commonly leach 100-1800 µg/L of total organic carbon (TOC) as determined by temperature, stagnation, surface area to volume ratio, pipe brand and age [56,170,172]. These levels of carbon, are far above the commonly cited threshold of 100 µg/L suggested to spur microbial growth in potable water main distribution systems [173]. However, the proportion of this released organic carbon that is assimilable is not clear. Many studies have demonstrated that some PEX pipes increase biofilm growth [59,140,147] and OP growth [59,140] relative to copper and iron. Unfortunately, it is unclear how general these effects are because the formulation of PEX used (e.g., PEX-b) varies from one manufacturer to another [170,172] and is typically proprietary and thus not cited in the available literature [59,140,147]. An experiment in the Netherlands using small-scale recirculating water heater systems (eight gallon tanks) connected to copper or PEX pipes (19.4 ft) attributed over three times higher *Legionella* bulk water levels in PEX pipe systems as compared to copper pipe systems although the authors did not determine if the difference was due to copper antimicrobial effects or leached organic carbon growth-promotion [140].

PVC pipes can leach 60–50,000 µg/L of TOC under typical water use conditions [50,56,174], of which roughly 50% was estimated to be assimilable [42]. Other studies indicate that PVC can promote biofilm growth [175,176] and proliferation of OPs compared to copper, lined cement, iron, and stainless steel [111,177–179]. When copper, glass, PEX, and PVC were used as materials in a biofilm
apparatus simulating premise plumbing, PVC and PEX materials maintained the highest Legionella growth potential in remineralized reverse osmosis water [178]. Other studies have drawn similar conclusions for other OPs compared to copper [111,128,148,149].

2.4. Iron Release from Pipes

Iron pipes may provide important niches and nutrients for OP growth. Antiquated cast iron, galvanized iron, and steel pipes in service lines and home plumbing can leach iron to water in a range of 0.2–18 mg/L dependent on factors including water chemistry, stagnation, surface area to volume ratio, and historical corrosion control [180,181]. Iron can also accumulate in loose deposit or biofilms and some studies have suggested that such locations are hotspots for growth of Legionella and other pathogens [40,182]. Studies examining M. avium have found that galvanized steel supported more growth than copper, PVC, and stainless steel [111,149].

Iron is an important nutrient for microorganisms involved in oxygen transfer, protein synthesis, and other essential metabolism [183] and some studies have shown that the presence of iron contributes to OP growth. Bench-scale studies have demonstrated that iron concentrations of up to 1 mg/L could enhance L. pneumophila growth in tap water while high concentrations (10, 100 mg/L) of iron produced toxic effects on L. pneumophila [184]. During the Legionnaires’ Disease outbreak in Flint, MI, our research found that the median iron concentration was 0.11 mg/L in cold water samples during the outbreak, but the outbreak’s end coincided with a water switch, dropping median iron in cold water samples down to less than 0.01 mg/L. Other field studies have observed similar positive correlations between L. pneumophila levels and iron concentrations [15,185]. In a simulated household drinking water system with no chlorine, van der Lugt et al. [186] observed that colonization of stainless steel faucets by Legionella was enhanced in the presence of 0.09 mg/L cast iron rust. It is important to note that in any study employing chlorine, iron pipe corrosion will remove the chlorine, confounding simplistic attribution of the higher Legionella to either iron or chlorine [26,187,188]. One study specifically examined if iron addition increased L. pneumophila growth without any chlorine present, and showed that it did so in one water with naturally low iron, but had no effect in another water with relatively high ambient iron [187].

2.5. Zinc, Aluminum, Magnesium Plumbing Materials

Pipes and plumbing devices can be composed of other metals that might affect the growth of OPs, but their impacts are largely unexplored. Zinc is present in source waters in concentrations ranging from <0.011 to 0.04 mg/L [189,190] and is normally below 0.1 mg/L in finished water [191]. Zinc concentrations at the tap are largely driven by its addition in corrosion inhibitors, or release from brass fixtures and galvanized pipes [190–192], and concentrations can reach 5 mg/L or higher [193,194]. Analogous to copper, zinc is an essential nutrient for microbial growth [195–200]. Zinc addition has been shown to increase L. pneumophila and P. aeruginosa growth in culture media [201], and high soluble zinc has been correlated with NTM [202].

Zinc can be toxic to microorganisms [196,203–206], but is believed to have limited biocidal activity compared to other metals [207], especially as it is below the US EPA Secondary Drinking Water Regulation limit of 5 mg/L [207] and Chinese Standard for Drinking Water Quality of 1 mg/L [208]. Inhibitory concentrations of zinc for Ops such as Pseudomonas spp., P. aeruginosa, and Aspergillus niger range from 13 to 650 mg/L in nutrient broth [204–206]. While this is a relatively high concentration range, Zhang et al. [180] demonstrated that galvanized iron pipes can release zinc to these levels in the presence of nitrifying bacteria. Furthermore, the biocidal activity of zinc or any other trace metal in premise plumbing will be controlled by the same chemistry factors including pH, hardness and NOM mentioned previously for copper.

Aluminum or magnesium rods are also commonly present as sacrificial anodes in water heaters (Figure 4), elevating Al$^{3+}$ or Mg$^{2+}$ levels in the water. Mg$^{2+}$ is known to be an essential nutrient for Legionella [201], whereas no such criteria have been established for Al$^{3+}$. More research is
needed to determine whether these additional trace metals encourage or discourage OP growth in plumbing systems.

3. Indirect Effects of Pipe Material on Pathogen Growth

3.1. Pipe Material Effect on Disinfectant Availability

Pipe material is a key factor affecting disinfectant decay in potable water systems. Maintaining relatively high levels of disinfectant residual is important to OP control because OPs are 20–600x more disinfectant resistant than the common indicator microorganisms such as E. coli [24] and are further protected in biofilms or host organisms [209–214]. Plastic pipe materials are generally non-reactive with chlorine and chloramine in terms of maintaining disinfectant residual levels, even though chlorine does sometimes slowly react with and degrade certain types of PEX and polyethylene pipe [44–49,51,215]. On the other hand, iron pipes have an extremely high disinfectant demand, as free chlorine cannot co-exist in equilibrium with ferrous or zero valent iron [44,46–48]. While chloramine is relatively non-reactive, iron oxide scale and associated nitrifying biofilms can cause relatively rapid monochloramine decay [216,217]. The reactivity of copper pipes and copper oxides is typically between plastics and iron and chemically catalyzes both chlorine and chloramine degradation [43,54,156,218–220]. Higher pH and the existence of phosphate can help maintain disinfectant residual levels in both iron and copper pipes [26,54].

3.2. Effect of Metallic Plumbing Materials on Nutrient Availability via Autotrophic Carbon Fixation

Although metallic plumbing does not leach assimilable organic carbon directly to water, certain metals can indirectly help OPs overcome carbon limitations by facilitating the growth of autotrophic microorganisms. Specifically, metallic pipes can encourage growth of hydrogen-oxidizing, ammonia-oxidizing, and ferrous-oxidizing autotrophic bacteria that fix inorganic carbon into new biomass [66,221].

3.2.1. Hydrogen Oxidizing Bacteria

The corrosion of iron pipes and the galvanic corrosion of aluminum or magnesium sacrificial anodes protecting steel water heaters can evolve hydrogen gas, which is a strong electron donor for autotrophs [60,61,110,221]. Ishizaki et al. [222] indicated that hydrogen-oxidizing bacteria, Alcaligenes eutrophus, could fix 2300 µg C/mmol H₂ in biomass in closed circuit cultivation system at gas pressure slightly higher than atmosphere, which could practically translate into production of up to 80 µg/L organic carbon biomass per day in an 80-gallon water heater equipped with a magnesium anode [223]. A study by Dai et al. [224] of an experimental water heater plumbing rig at 39, 42, and 51 °C confirmed elevated levels of functional genes associated with hydrogen metabolism, demonstrating that hydrogen-oxidizing bacteria were able to proliferate in water heaters.

3.2.2. Autotrophic Ammonia and Iron Oxidizing Bacteria

Iron and copper can catalyze the conversion of chloramine disinfectant to free ammonia, which can then serve as a substrate for autotrophic ammonia oxidizing bacteria. Ammonia-oxidizing bacteria can fix substantial amounts of organic carbon into the system, specifically 21 to 240 µg C/mg NH₃-N based on experimental growth yield values of pure or mixed cultures [225]. Ferrous iron, released as a natural by-product of iron corrosion, can also fix an average of 26 µg C/mg Fe²⁺ under circumneutral condition measured in bioreactors [226].

3.2.3. Copper Deposition Corrosion Accelerating H₂ Evolution

Although copper cannot corrode with evolution of H₂ gas, cupric ions in water can plate onto the less noble metals (zinc, aluminum, iron and magnesium) via deposition corrosion. This copper coating can dramatically accelerate corrosion of less noble metals and indirectly stimulate evolution of
hydrogen (H\textsubscript{2}) gas (Figure 4) [66,222,227,228]. A study using a combination of bench- and pilot-scale hot water system experiments demonstrated these effects [222].

**Figure 4.** Water heater material interactions create multiple niches suitable for bacterial and opportunistic pathogen (OP) growth. Deposition of copper onto less noble metals (e.g., a water heater anode) can result in dramatically accelerated corrosion and release dissolved H\textsubscript{2} gas, which is an electron donor for autotrophs. If the anode rod consists of magnesium, then the pH will become elevated as well. Figure adapted from Brazeau et al. [229].

### 3.3. Pipe Scaling Effects

Scaling caused by pipe corrosion or higher pH can increase pipe surface roughness, which is known to enhance biofilm colonization and overall growth, creating an ideal environment for OP establishment and proliferation [112]. One study showed that copper coupons in a biofilm reactor formed extensive scales and promoted seven-fold more biofilm biomass than PVC pipes after three months of incubation [230]. Aged metal pipes may form very thick scales characterized by corrosion tubercles and extensive networks of pores [60,231–233], providing an area for not only additional biofilm growth, but also distinct microenvironments [233,234] with pH as low as 2.0 or as high as 10 [235].

### 4. Influence of Plumbing System Design, Configuration and Operation

All of the direct and indirect interactions described in previous sections are further influenced by the specific premise plumbing design, configuration, and operation. Flow rate, water stagnation, temperature profile, secondary disinfectant concentration, and nutrient availability can all interact to create hot spots for OPs growth in buildings.

#### 4.1. Water Stagnation

Water age is defined as the time it takes water to move from one point to another in the system, which may influence OP growth through a variety of mechanisms. This includes the time from when...
it is freshly produced at the treatment plant and travels to the service line, as well as the time from when it first enters the building’s plumbing to the point of use [71]. High water age in buildings is increased by: (1) existence of dead ends/legs and stagnation in plumbing systems [182,236]; (2) use of low flow devices or presence of large storage tanks such as those used for solar water heating or onsite rainwater collection [39]; and (3) using low volumes of water in a building or at a particular outlet [192]. Stagnation and infrequent water use may concentrate and enhance release of organic matter in water in plastic pipes and metals in metallic pipes [181,237–240]. Zhang et al. [241] found a four-fold increase in bulk water TOC in unplasticized PVC pipes between 24 h and 72 h of stagnation. Fixtures in a green building with the fewest water use events (most stagnation) also had greater organic carbon, bacteria counts, and heavy metal (Zn, Fe, Pb) concentrations [192,242].

Stagnation and high water age also increases the likelihood and rate of disinfectant decay. High consumption of chlorine and chloramine during stagnant periods of 24–72 h have been observed for synthetic pipes (0.4 and 0.6 mg/L of chlorine loss, respectively), and stagnant periods of 2–8 h in metallic pipes (3 and 4 mg/L chlorine loss, 1.5 and 3.5 mg/L chloramine loss, respectively) [54,241]. In a green building study, six-hour stagnation almost fully eliminated monochloramine (>99%) within pipes [71].

Such water quality changes have been related to increased levels of OPs in premise plumbing systems [39,243–245]. In a field sampling study of main water distribution system, 120 water samples were taken throughout a drinking water distribution system. Only four samples were positive for cultivable L. pneumophila and all four samples were taken from dead end points at the end of streets with no chlorine residual remaining [246]. Another field study identified their most frequently Legionella positive sites as being located at the end of the distribution system and having the highest turbidity, iron, TOC, and water age, as well as the lowest flow [247]. The association between OPs and stagnation has created interest in strategies to reduce building water stagnation effects such as removing dead-legs, flushing, maintaining the hot water system, and shock disinfection [248–251]. The effectiveness of these strategies should be evaluated within the context of the specific pipe materials that are present.

4.2. Hot Water Recirculation Lines

Some plumbing codes require or suggest the use of recirculating hot water lines for water/energy conservation, convenience and comfort [1,252–254]. In these systems, water is circulated continuously between the water heater and the point of use, preventing cooling of the distal lines and allowing for nearly instant delivery of hot water at the point of use [255]. There are many important differences between hot water recirculating systems and conventional systems, which are stagnant during periods of disuse that can affect OP growth. The constantly flowing water can deliver more nutrients to biofilm and hypothetically increase OP growth [230]. On the other hand, continuous flow can deliver more disinfectants and more hot water, which are critical control measures for OPs [256,257]. The net effect depends on which of these factors is dominant.

Continuously recirculating water could also increase release of metals, increase deposition corrosion of anodes by constantly recirculating water through copper pipe, and result in greater accumulation of sediments and H₂ gas. One study showed that recirculating systems with copper piping had 3–13 times more aluminum and copper, 4–6 times more hydrogen in effluent water, and 9% more aluminum anode weight loss, compared with standard (non-recirculating) systems [222]. Recirculating systems can also accumulate 3–20 times more sediments [222] arising from corrosion of metallic pipe material and the anode rods [157,232–234,258]. These sediments, which also collect at the bottom of hot water tanks, may serve as an important growth niche within warm regions of hot water tanks where influent cold water depresses temperatures, and there are also relatively low levels of disinfectant and high levels of nutrients for Legionella, heterotrophs, and host organisms [17,259].
4.3. Pipe Aging

New plastic and copper pipes behave differently than older pipes. Specifically, corrosion and release of metals is strongly influenced by pipe age, with corrosion rates and metal release tending to decrease as thicker and more passivating pipe scales form. Aging can dramatically reduce levels of metal leaching from copper and other pipes [157,260,261]. The rate of aging, and whether it decreases release of pipe constituents at all, is highly affected by water chemistry and water use patterns [157]. Likewise, leaching of organics from plastic pipe may attenuate 50% to >99% after aging for a period of a few weeks with hot water exposure [51,170], but in other cases has been sustained for months [262] or even over a year [263]. Pipe aging is an important factor to consider when comparing PEX to copper’s capacity for Legionella growth. One study showed that the Legionella numbers in bulk water of both PEX and copper pipes in a simulated warm water system were the same after two years [140]. We speculate that one possible cause for this convergence is that, as plastic pipes age, organic carbon migration to water decreases, whereas levels of antimicrobial copper released from copper pipe also tends to decrease. Hence, in some situations, it is expected that in very old copper and plastic pipe systems there would be little difference between these pipe materials.

4.4. Possible Mixed Material Interactions

Building plumbing is typically comprised of multiple pipe materials, either in the original design or after partial retrofits or renovations. It is anticipated that there are sometimes synergistic and other times antagonistic interactions between pipe materials that would influence growth of OPs. Copper deposition accelerating the evolution of H$_2$ from aluminum, zinc, magnesium and iron corrosion, as discussed in Section 3.2.3, is an important exemplar. Copper is also known to catalyze degradation of plastic pipes [264–268], and the presence of copper pipe upstream of plastic pipe might enhance organic carbon release [268], surface roughness for biofilm growth [264], and perhaps even disinfectant consumption due to copper in the scale. Iron pipes upstream of copper may produce mixed Fe-Cu oxides, which can be extraordinary catalysts for free chlorine decay [269]. Similarly, copper released upstream of iron pipes could increase iron release [270]. Any galvanic coupling between two metals in plumbing materials (copper/brass-lead [271,272], copper/iron [270,273,274] iron/zinc [275,276], copper/aluminum [277,278], copper/zinc [271,279], copper/magnesium [280]) has the potential to enhance corrosion and cause changes to water quality parameters relevant to corrosion and OP growth [235,281], dissolved oxygen (DO) [273], metal concentrations [271,272], and disinfectant residual concentration. These reactions also create microenvironments of very high or very low pH [235,238]. Given that in the 2017 American Housing Survey 10% of households that reported any home improvement projects also reported adding or replacing an interior water pipe [282], understanding the effects of mixing pipe materials during renovation appears to be a valuable research area as antiquated premise plumbing is increasingly replaced.

5. Mediating Role of Microbiome and other Microbiological Considerations

5.1. The Role of Pipe Material in Shaping the Premise Plumbing Microbiome and Resident Amoeba Host Organisms

Interactions between OPs and the microbial communities surrounding them are key to OP proliferation and are likely influenced by pipe materials. OPs can be parasitic to free-living amoebae that first prey upon them in drinking water biofilms, before they reproduce inside and eventually kill the host organism [24]. In fact, there is some doubt that Legionella actually reproduces significantly in drinking water outside of an amoeba host [283]. Amoebae can also protect OPs from disinfectants and provide access to nutrients. For example, Legionella exclusively use amino acids, which are abundant in amoeba vacuoles, as a carbon source [210–214,284,285]. Thus, although poorly studied, any factor altering growth of key host amoebae (including Acanthamoeba, Vermamoeba, and Naegleria) is expected to indirectly affect growth of OPs, including L. pneumophila, P. aeruginosa,
and NTM [122,210–214,225,257,286,287]. In one experiment, copper coupons were found to host more *Acanthamoeba polyphaga* than PVC coupons [288], possibly because copper hosts less diverse eukaryotic communities [64,289] and limits competition for *A. polyphaga*. As a result, *L. pneumophila* grew and shed to the bulk waters in higher numbers on these copper coupons than on PVC coupons if co-inoculated with *A. polyphaga* [122].

Interbacterial interactions may also influence the growth of OPs. Broadly speaking, OPs benefit from the biofilm community through access to nutrients and protection from disinfectants [24,35,36,290]. Some studies have identified correlations between specific taxa and OPs in premise plumbing [291], cooling towers [292] and drinking water distribution systems [293]. However, the significance of these correlations to premise plumbing material selection is not well understood, as most studies examining differences in bacterial communities focus on very broad measures of community structure [48,59,64,216,289,294–296]. Certain waterborne bacteria are known to produce toxins that inhibit *L. pneumophila* growth [216,297] or exude other compounds that have secondary bacteriostatic effects on *Legionella* [298]. Intra-bacterial inhibition also may be mediated through amoebae by reducing host uptake [299,300] or killing the host population [134,301,302]. More research is needed to elucidate how the broad ecological differences resulting from pipe material influence these interactions. Integration of metagenomic or meta-transcriptomic analyses targeting the production of bacteriocins or other toxins with known effects on OPs could elucidate the ecological effects of taxonomic shifts resulting from pipe material. Interrupting OP-amoea endosymbiosis through the enrichment of preferential non-OP amoeba prey [299,300] has been suggested as a probiotic means of controlling OPs [303], and pipe material could be explored as a means of enrichment of these taxa.

### 5.2. Variation in Copper Tolerance Among Species and Strains

Strain-to-strain differences in intrinsic tolerance of copper, acclimation to copper concentrations with time through induction of the appropriate genes, or acquisition of copper resistance via mutation or horizontal gene transfer in premise plumbing might explain some of the discrepancies in variable outcomes of copper on OPs (Table 2). *Legionella* [155] and other OPs [58] may acclimate to high copper levels through the expression of copper detoxification or efflux systems. Bedard et al. [155] reported four-fold differences in the copper tolerance of environmentally-isolated *L. pneumophila* strains, noting that more resistant strains showed increased copper ATPase *copA* expression, speculating that their increased tolerance may also be a result of higher biofilm production. Strikingly, Williams et al. [58] showed that, during exposure to 95 mg/L of copper over 6 h in liquid culture, culturable *A. baumannii* levels (CFU/mL) could increase by 2-logs or decrease by 2-logs, depending on the strain. The authors identified putative copper detoxification and efflux systems within the genome of the most resistant isolate and identified specific genes that were upregulated in response to copper exposure. However, a majority of the less tolerant strains tested also possessed these genes, leading the authors to suggest that further definition of the proteins involved in copper resistance is required. One recent study showed two environmentally-isolated *Legionella* strains reduced by less than one log in culturability, even after two weeks of exposure to 5 mg/L copper, which the authors attributed to adaptation to the high levels of copper (average 0.48 mg/L) in the hot water system from which these isolates were collected [154]. A profile of *Fusarium* isolates revealed that tap water isolates were more copper-tolerant than soil isolates [303]. *P. aeruginosa* isolates isolated from a hospital with copper plumbing exhibited only slightly limited growth in the presence of 0.15 mg/L copper [129]. All of these strains were found to harbor GI-7, a mobile genetic element that confers copper resistance and that has also been identified in a *P. aeruginosa* strain associated with hospital outbreaks [304]. Limited data suggest that *A. baumannii* and mycobacteria are more difficult to inactivate with copper than other OPs, while *P. aeruginosa* is more readily inactivated [91,92,98,108,109]. *L. pneumophila* has been found both at the more resistant [98] and less resistant [91,108,109] ends of this spectrum. The wide variability among OPs and even strains of OPs in their intrinsic tolerance of copper, ability to
acquire genetic resistance, and ability to acclimate to elevated levels of copper makes it difficult to precisely predict the efficacy of copper and other antimicrobials for OP control.

5.3. Confounding Effects of VBNC Bacteria

The discovery of VBNC bacteria has complicated prior understanding for all OP control strategies, including copper. Virtually all prior work relied on culture methods to determine copper’s efficacy for killing OPs [62,63,83,91,92,98,108,109,120,137,153], but some microbes rendered not culturable might remain viable and still infect host amoebae or humans [74,76,305–307]. The existence of VBNC pathogens in premise plumbing has been demonstrated by comparing culture-based numbers with those enumerated via fluorescence (e.g., live/dead) and molecular-based (e.g., quantitative polymerase chain reaction) monitoring methods [308].

Bench-scale studies examining copper’s antimicrobial efficacy have found discrepancies between culture-based and molecular-based numbers of *L. pneumophila* [121,122] that are also suggestive of a copper-induced VBNC state. Similar discrepancies have been noted for *P. aeruginosa*, *Stenotrophomonas maltophilia*, and *M. avium* [104,109,127,132,133]. Evidence of copper-induced VBNC activity is particularly strong in the case of *P. aeruginosa*, where one study applied multiple non-culture-based measures of viability [127,132]. Furthermore, VBNC *P. aeruginosa* have been shown to partially recover infectivity after removal of copper from solution [132,133]. To understand how VBNC bacteria contribute to OP infections, additional studies are needed to delineate the premise plumbing conditions more precisely that induce VBNC status and to confirm the range of functionality maintained in this state. A primary challenge in achieving this is that there are currently no reliable methods for confidently enumerating VBNC bacteria.

5.4. Virulence

The premise plumbing environment exhibits several features that could possibly contribute to the virulence of resident OPs. Wargo [38] describes features of drinking water plumbing that could prime OPs to infect cystic fibrosis patients, although the interactions described in this review could also pose risk to otherwise immunocompromised individuals. Such features that are relevant to pipe material include [38]:

- Elevated copper levels, selecting for resistance to copper overload within macrophage phagosomes, a component of the innate immune response [309].
- Elevated iron levels, influencing interactions between iron homeostasis and virulence.
- Exposure to lipids, which are generally not well removed by drinking water treatment, priming OPs for lipid-rich environments within hosts. Accumulation of phospholipid fatty acids has been shown to be greater in the biofilms of polyethylene pipes than copper pipes, though these lipids were putatively associated with bacteria [310].
- Low DO levels, selecting for OPs capable of survival in low DO regions of the biofilm in infected host tissue.
- Exposure to eukaryotic predation, selecting for resistance to the host’s immune response (e.g., lung macrophages) or enhanced virulence.

Some studies suggest that the above types of interaction may increase the pathogenic potential of premise plumbing-associated OPs specifically. Copper resistance is important to mammalian host infection for *P. aeruginosa* [311] and *A. baumannii* [312,313], and other evidence suggests that exposure to copper in aquatic environments selects for greater copper resistance among certain OPs [129,303,304]. Copper and other divalent metals may also play a role in nutrient acquisition and pathogenesis even after infecting hosts [314].

The effects of iron exposure on OPs are not as apparent. *L. pneumophila* serogroup 1 grown in medium that was iron limited (0.017–0.056 mg/L) has been shown to lose its virulence [315], indicating that limiting adequate concentrations of iron could not only decrease the presence of
Legionella but also the likelihood of human infection. Iron also plays a role in modulating various behaviors, including modulating virulence factor production in \textit{P. aeruginosa} and \textit{A. baumannii} [316–321], but it is unclear what effects exposure to iron have on virulence in the premise plumbing environment. This subject is largely unexplored and more research is needed to determine the overall effects of the premise plumbing environment on OP virulence.

5.5. Antibiotic Resistance and Tolerance

Copper, among other heavy metals has been shown to exert selection pressure, leading to enhanced survival of antibiotic resistant bacteria. In fact, heavy-metal-associated co-selection and cross-selection has been proposed to be as much of a concern for environmental propagation of antibiotic resistance as antibiotics themselves [322]. Increases in antibiotic resistance genes at the community scale have been identified after long-term copper exposure in soil [323–326], sediment [327], and drinking water [327]. Bench-scale tests using bacterial isolates from biofilters [328] and wastewater [329] inoculated into growth media have shown that a selective or inductive effect of copper can take places within hours. However, these studies were performed with copper concentrations 5–77 times greater than the 1.3 mg/L US EPA copper action level and similarly in exceedance of the Chinese Standard for Drinking Water Quality of 1 mg/L [209] and WHO Guideline for Drinking-Water Quality of 2 mg/L [82]. Thus, these concentrations may not be representative of potable water systems. One study examining antibiotic resistant and sensitive strains of \textit{Staphylococcus aureus} showed that the more antibiotic resistant strain survived longer in a copper container [90]. As discussed above, copper may also better support \textit{Acanthamoeba} than other materials, while in one study \textit{L. pneumophila} grown within \textit{A. polyphaga} demonstrated increased tolerance to all antibiotics tested (rifampin, ciprofloxacin, and erythromycin) compared to those grown in culture media [330]. The role of copper plumbing and other pipe materials in these emerging areas of research is worthy of further investigation.

There is more limited evidence that the presence of iron may also induce or select for antibiotic resistance, as observed for \textit{P. aeruginosa} using iron-amended growth media [330] and the gut microbiomes of mice supplied with iron-amended water [331]. The latter case, while using an iron concentration more than 25 times the EU drinking water standard of 0.2 mg/L [332] and 16 times both the US EPA National Secondary Drinking Water Standard and Chinese Standard for Drinking Water Quality of 0.3 mg/L, may be of particular concern, as it suggests that pipe corrosion products have the potential to select for antibiotic resistance inside the infected host organism.

6. Conclusions

Premise plumbing is a complex, temporally dynamic, and spatially diverse environment that is strongly influenced by pipe materials. Virtually all pipe materials have known benefits and/or detriments for OP growth. Plumbing materials are an important driver of the chemical and biological water quality parameters that influence the control of OPs and there are no silver (copper or plastic) bullets that will uniformly inhibit the growth of \textit{Legionella} and other OPs under all circumstances.

Synthetic plastic pipe materials vary between type and manufacturer. They can act as a supply of organic carbon for the growth of microorganisms, but exert a lower chlorine demand and tend to form fewer scales that could provide more surface area for biofilm growth. Iron pipes supply nutrients for growth, exhibit a high disinfectant demand, produce hydrogen and other nutrients through corrosion, and tend to form thick scales with extremely high surface areas. While they may no longer be used in new construction, even short sections of pipe can affect an entire downstream premise plumbing distribution system. Stainless steel has few known effects on water quality, and correspondingly, OP control, perhaps because it is the least studied and is less commonly used as a result of its high cost. Copper pipes are known for their antimicrobial ability, but this is inconsistently realized in practice, and in some cases they seem to encourage OP growth relative to other pipes. Premise plumbing materials have a role to play in preventing OP infections and, at a minimum, should be examined more
closely for their propensity to inhibit or stimulate OP proliferation during outbreak investigations. Research is needed to better define:

• Both the intra-species and inter-species variation of copper resistance amongst OPs, as well as environmental drivers of this variation.
• Effects of copper pipes on OPs in a more holistic sense, with identification of real-world conditions that are drivers for discrepancies in copper’s antimicrobial capacity.
• Copper’s possible micronutrient activity in OPs within premise plumbing contexts, including threshold concentrations required for various physiological functions, as well as physicochemical and ecological factors that influence those thresholds.
• The disease risk that VBNC OPs pose and conditions under which copper and other antimicrobials induce VBNC status in premise plumbing OPs
• The inhibitory action of trace metals on OP growth in premise plumbing, as well as growth requirements for other trace elements exhibited by OPs in premise plumbing.
• Potential mediating effects of the wider microbial community composition resulting from pipe material on OPs.
• Effect of mixed pipe materials on physicochemical parameters of bulk water and OP growth.
• The effects of plumbing materials on OP antibiotic resistance and virulence.
• The impact of stagnation, velocity, sediments, corrosion control, and consumer water use patterns on all of the above.

An improved understanding will provide actionable advice for multiple stakeholders. In addition to the obvious direct use of the results in the construction industry and by building water quality managers, water utilities can benefit from improved understanding of how the interplay of premise plumbing pipe materials with disinfectants, nutrients and corrosion control can be harnessed to reduce disease incidence.

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References
1. National Academies of Sciences, Engineering, and Medicine. Management of Legionella in Water Systems; The National Academies Press: Washington, DC, USA, 2019; ISBN 978-0-309-49947-7.
2. Ashbolt, N.J. Microbial Contamination of Drinking Water and Human Health from Community Water Systems. Curr. Environ. Health Rep. 2015, 2, 95. [CrossRef] [PubMed]
3. United States Occupational Safety & Health Administration Legionnaires’ Disease: Facts and Frequently Asked Questions. Available online: https://www.osha.gov/dts/osta/otm/legionnaires/faq.html (accessed on 1 July 2020).
4. United States Occupational Safety & Health Administration Legionnaires’ Disease. Available online: https://www.osha.gov/dts/osta/otm_iii/otm_iii_7.html (accessed on 1 July 2020).
5. CDC Legionella (Legionnaires’ Disease and Pontiac Fever) Surveillance and Reporting. Available online: http://www.cdc.gov/legionella/surv-reporting.html (accessed on 1 July 2020).
6. Singh, R.; Hamilton, K.A.; Rashiduzzaman, M.; Yang, Z.; Kar, S.; Fasnacht, A.; Masters, S.V.; Gurian, P.L. Managing Water Quality in Premise Plumbing: Subject Matter Experts’ Perspectives and a Systematic Review of Guidance Documents. *Water* 2020, 12, 347. [CrossRef]

7. McCoy, W.F. Preventing Legionellosis; IWA Publishing: London, UK, 2005; ISBN 1843390949.

8. Falkinham, J.; Pruden, A.; Edwards, M. Opportunistic Premise Plumbing Pathogens: Increasingly Important Pathogens in Drinking Water. *Pathogens* 2015, 4, 373. [CrossRef]

9. Centers for Disease Control Legionella (Legionnaires’ Disease and Pontiac Fever). Available online: https://www.cdc.gov/legionella/index.html (accessed on 1 July 2020).

10. Gomez-Valero, L.; Rusniok, C.; Rolando, M.; Neou, M.; Dervins-Ravault, D.; Demirtas, J.; Rouy, Z.; Moore, R.J.; Chen, H.; Petty, N.K.; et al. Comparative analyses of Legionella species identifies genetic features of strains causing Legionnaires’ disease. *Genome Biol.* 2014, 15, 505. [CrossRef]

11. Association of Water Technologies Legionella 2019: A Position Statement and Guidance Document. Available online: https://www.awt.org/pub/035C2942-03BE-3BFF-08C3-4C686FB7395C (accessed on 15 November 2020).

12. Allen, M.J.; Edberg, S.C.; Clancy, J.L.; Hrudey, S.E. Drinking water microbial myths. *Crit. Rev. Microbiol.* 2015, 41, 366–373. [CrossRef]

13. Barrabeig, I.; Rovira, A.; Garcia, M.; Oliva, J.M.; Vilamala, A.; Ferrer, M.D.; SabrìÀ, M.; DomíNguez, A. Outbreak of Legionnaires’ disease associated with a supermarket mist machine. *Epidemiol. Infect.* 2010, 138, 1823–1828. [CrossRef] [PubMed]

14. Mahoney, F.J.; Hoge, C.W.; Farley, T.A.; Barbaree, J.M.; Breiman, R.F.; Benson, R.F.; McFarland, L.M. Communitywide Outbreak of Legionnaires’ Disease Associated with a Grocery Store Mist Machine. *J. Infect. Dis.* 1992, 165, 736–739. [CrossRef] [PubMed]

15. Stout, J.E.; Yu, V.L.; Yee, Y.C.; Vaccarello, S.; Diven, W.; Lee, T.C. Legionella pneumophila in residential water supplies: Environmental surveillance with clinical assessment for Legionnaires’ disease. *Epidemiol. Infect.* 1992, 109, 49–57.

16. Stout, J.E.; Victor, L.Y.; Muraca, P. Isolation of Legionella pneumophila from the cold water of hospital ice machines: Implications for origin and transmission of the organism. *Infect. Control Hosp. Epidemiol.* 1985, 6, 141–146. [CrossRef]

17. Rhoads, W.J.; Bradley, T.N.; Mantha, A.; Buttling, L.; Keane, T.; Pruden, A.; Edwards, M.A. Residential water heater cleaning and occurrence of Legionella in Flint, MI. *Water Res.* 2020, 115439. [CrossRef]

18. Gobin, I.; Newton, P.R.; Hartland, E.L.; Newton, H.J. Infections caused by nonpneumophila species of Legionella. *Rev. Med. Microbiol.* 2009, 20, 1–11. [CrossRef]

19. Joseph, C.A. Legionnaires’ disease in Europe 2000–2002. *Epidemiol. Infect.* 2004, 132, 417–424. [CrossRef] [PubMed]

20. Stout, J.E.; Yu, V.L.; Muraca, P.; Joly, J.; Troup, N.; Tompkins, L.S. Potable water as a cause of sporadic cases of community-acquired legionnaires’ disease. *N. Engl. J. Med.* 1992, 326, 151–155. [CrossRef]

21. Fliermans, C.B.; Cherry, W.B.; Orrison, L.H.; Smith, S.J.; Tison, D.L.; Pope, D.H. Ecological distribution of *Legionella pneumophila*. *Appl. Environ. Microbiol.* 1981, 41, 9–16. [CrossRef] [PubMed]

22. Lee, G.F.; Jones-Lee, A. Public Health Significance of Waterborne Pathogens in Domestic Water Supplies and Reclaimed Water. In *Report to State of California Environmental Protection Agency Comparative Risk Project; California Environmental Protection Agency*; Berkeley, CA, USA, 1993.

23. Borella, P.; Montagna, M.T.; Romano-Spica, V.; Stampi, S.; Stancanelli, G.; Triassi, M.; Neglia, R.; Marchesi, L.; Fantuzzi, G.; Tato, D.; et al. Legionella Infection Risk from Domestic Hot Water. *Emerg. Infect. Dis.* 2004, 10, 457–464. [CrossRef] [PubMed]

24. Falkinham III, J.O. Common Features of Opportunistic Premise Plumbing Pathogens. *Int. J. Environ. Res. Public Health* 2015, 12, 4533–4545. [CrossRef] [PubMed]

25. Parr, A.; Whitney, E.A.; Berkelman, R.L. Legionellosis on the Rise: A Review of Guidelines for Prevention in the United States. *J. Public Health Manag. Pract.* 2015, 21, E17–E26. [CrossRef]

26. Rhoads, W.J.; Garner, E.; Ji, P.; Zhu, N.; Parks, J.; Schwake, D.O.; Pruden, A.; Edwards, M.A. Distribution System Operational Deficiencies Coincide with Reported Legionnaires’ Disease Clusters in Flint, Michigan. *Environ. Sci. Technol.* 2017, 51, 11986–11995. [CrossRef]

27. Mercante, J.W.; Winchell, J.M. Current and Emerging Legionella Diagnostics for Laboratory and Outbreak Investigations. *Clin. Microbiol. Rev.* 2015, 28, 95–133. [CrossRef]
28. Garrison, L.E. Vital Signs: Deficiencies in Environmental Control Identified in Outbreaks of Legionnaires’ Disease—North America, 2000–2014. MMWR. Morb. Mortal. Wkly. Rep. 2016, 65, 576–584. [CrossRef]

29. Pierre, D.; Baron, J.L.; Ma, X.; Sidari, F.P.; Wagener, M.M.; Stout, J.E. Water Quality as a Predictor of Legionella Positivity of Building Water Systems. Pathogens 2019, 8, 295. [CrossRef]

30. Perrin, Y.; Bouchon, D.; Delafont, V.; Moulin, L.; Héchard, Y. Microbiome of drinking water: A full-scale spatio-temporal study to monitor water quality in the Paris distribution system. Water Res. 2019, 149, 375–385. [CrossRef] [PubMed]

32. Pinto, A.J.; Xi, C.; Raskin, L. Bacterial community structure in the drinking water microbiome is governed by filtration processes. Environ. Sci. Technol. 2012, 46, 8851–8859. [CrossRef] [PubMed]

33. Wang, H.; Edwards, M.A.; Pruden, A. Distribution System Water Quality Affects Responses of Opportunistic Pathogen Gene Markers in Household Water Heaters. Environ. Sci. Technol. 2013, 47, 10117–10128. [CrossRef]

34. Baron, J.L.; Vikram, A.; Duda, S.; Stout, J.E.; Bibby, K. Shift in the Microbial Ecology of a Hospital Hot Water System following the Introduction of an On-Site Monochloramine Disinfection System. PLoS ONE 2014, 9. [CrossRef] [PubMed]

35. Schwering, M.; Song, J.; Louie, M.; Turner, R.J.; Ceri, H. Multi-species biofilms defined from drinking water microorganisms provide increased protection against chlorine disinfection. Biofouling 2013, 29, 917–928. [CrossRef] [PubMed]

37. Wang, H.; Edwards, M.A.; Pruden, A. Probiotic Approach to Pathogen Control in Premise Plumbing Systems? A Review. Environ. Sci. Technol. 2013, 47, 10117–10128. [CrossRef]

40. Liu, G.; Tao, Y.; Zhang, Y.; Lut, M.; Knibbe, W.-J.; van der Wielen, P.; Liu, W.; Medema, G.; van der Meer, W. Hotspots for selected metal elements and microbes accumulation and the corresponding water quality deterioration potential in an unchlorinated drinking water distribution system. Water Res. 2017, 124, 435–445. [CrossRef] [PubMed]

42. Neu, L.; Hammes, F. Feeding the Building Plumbing Microbiome: The Importance of Synthetic Polymeric Materials for Biofilm Formation and Management. Water 2020, 12, 1774. [CrossRef]
49. Lehtola, M.J.; Miettinen, I.T.; Lampola, T.; Hirvonen, A.; Vartiainen, T.; Martikainen, P.J. Pipeline materials modify the effectiveness of disinfectants in drinking water distribution systems. *Water Res.* **2005**, *39*, 1962–1971. [CrossRef]

50. Heim, T.H.; Dietrich, A.M. Sensory aspects and water quality impacts of chlorinated and chloraminated drinking water in contact with HDPE and cPVC pipe. *Water Res.* **2007**, *41*, 757–764. [CrossRef] [PubMed]

51. Mao, G.; Wang, Y.; Hammes, F. Short-term organic carbon migration from polymeric materials in contact with chlorinated drinking water. *Sci. Total Environ.* **2018**, *613*, 1220–1227. [CrossRef] [PubMed]

52. Durand, M.L.; Dietrich, A.M. Contributions of silane cross-linked PEX pipe to chemical/solvent odours in drinking water. *Water Sci. Technol.* **2007**, *55*, 153–160. [CrossRef] [PubMed]

53. Kim, H.; Koo, J.; Kim, S. A general framework of chlorine decay modeling at a pilot-scale water distribution system. *J. Water Supply Res. Technol. AQUA* **2015**, *64*, 543–557. [CrossRef]

54. Nguyen, C.; Elfland, C.; Edwards, M. Impact of advanced water conservation features and new copper pipe on rapid chloramine decay and microbial regrowth. *Water Res.* **2012**, *46*, 611–621. [CrossRef]

55. Westbrook, A.; Digiano, F.A. Rate of chloramine decay at pipe surfaces. *J. Am. Water Works Assoc.* **2009**, *101*, 59–70. [CrossRef]

56. Bucheli-Witschel, M.; Kötzsch, S.; Darr, S.; Widler, R.; Egli, T. A new method to assess the influence of material on microbiota composition and Legionella pneumophila in hot water plumbing systems. *Microbiome* **2017**, *5*, 130. [CrossRef]

57. Skjevrak, I.; Due, A.; Gjerstad, K.O.; Herikstad, H. Volatile organic components migrating from plastic pipes (HDPE, PEX and PVC) into drinking water. *Water Res.* **2003**, *37*, 1912–1920. [CrossRef]

58. Williams, C.L.; Neu, H.M.; Gilbreath, J.J.; Michel, S.L.J.; Zurawski, D.V.; Merrell, D.S. Copper Resistance of the Emerging Pathogen Acinetobacter baumannii. *Appl. Environ. Microbiol.* **2016**, *82*, 6174–6188. [CrossRef]

59. Proctor, C.R.; Dai, D.; Edwards, M.A.; Pruden, A. Interactive effects of temperature, organic carbon, and pipe material on microbiota composition and Legionella pneumophila in hot water plumbing systems. *Microbiome* **2017**, *5*, 130. [CrossRef]

60. Rushing, J.C.; McNeill, L.S.; Edwards, M. Some effects of aqueous silica on the corrosion of iron. *Water Res.* **2003**, *37*, 1080–1090. [CrossRef]

61. Niu, L.; Cheng, Y.F. Corrosion behavior of X-70 pipe steel in near-neutral pH solution. *Appl. Surf. Sci.* **2007**, *253*, 8626–8631. [CrossRef]

62. Lin, Y.-S.E.; Vidic, R.D.; Stout, J.E.; Yu, V.L. Individual and combined effects of copper and silver ions on inactivation of Legionella pneumophila. *Water Res.* **1996**, *30*, 1905–1913. [CrossRef]

63. Mathys, W.; Stanke, J.; Harmuth, M.; Junge-Mathys, E. Occurrence of Legionella in hot water systems of single-family residences in suburbs of two German cities with special reference to solar and district heating. *Int. J. Hyg. Environ. Health* **2008**, *211*, 179–185. [CrossRef] [PubMed]

64. Lu, J.; Buse, H.Y.; Gomez-Alvarez, V.; Struewing, I.; Santo Domingo, J.; Ashbolt, N.J. Impact of drinking water conditions and copper materials on downstream biofilm microbial communities and Legionella pneumophila colonization. *J. Appl. Microbiol.* **2014**, *117*, 905–918. [CrossRef] [PubMed]

65. Merritt, K.; Brown, S.A. Release of hexavalent chromium from corrosion of stainless steel and cobalt—Chromium alloys. *J. Biomed. Mater. Res.* **1995**, *29*, 627–633. [CrossRef] [PubMed]

66. Morton, S.C.; Zhang, Y.; Edwards, M.A. Implications of nutrient release from iron metal for microbial regrowth in water distribution systems. *Water Res.* **2005**, *39*, 2883–2892. [CrossRef]

67. Garner, E.; Zhu, N.; Strom, L.; Edwards, M.; Pruden, A. A human exposome framework for guiding risk management and holistic assessment of recycled water quality. *Environ. Sci. Water Res. Technol.* **2016**, *2*, 580–598. [CrossRef]

68. Anderson, L.E.; Krkošek, W.H.; Stoddart, A.K.; Trueman, B.F.; Gagnon, G.A. Lake Recovery Through Reduced Sulfate Deposition: A New Paradigm for Drinking Water Treatment. *Environ. Sci. Technol.* **2017**, *51*, 1414–1422. [CrossRef]

69. Monteith, D.T.; Stoddard, J.L.; Evans, C.D.; De Wit, H.A.; Forsius, M.; Høgåsen, T.; Wilander, A.; Skjelkvåle, B.L.; Jeffries, D.S.; Vuorenmaa, J.; et al. Dissolved organic carbon trends resulting from changes in atmospheric deposition chemistry. *Nature* **2007**, *450*, 537–540. [CrossRef]

70. European Environment Agency. Air Quality Standards. Available online: https://www.eea.europa.eu/themes/air/air-quality-concentrations/air-quality-standards (accessed on 28 October 2020).
71. Rhoads, W.J.; Pruden, A.; Edwards, M.A. Survey of green building water systems reveals elevated water age and water quality concerns. *Environ. Sci. Water Res. Technol.* 2015, 2, 164–173. [CrossRef]

72. United Nations. Transforming Our World: The 2030 Agenda for Sustainable Development. Available online: https://sustainabledevelopment.un.org/post2015/transformingourworld/publication (accessed on 15 November 2020).

73. Lee, J.; Kleczyk, E.; Bosch, D.J.; Dietrich, A.M.; Lohani, V.K.; Loganathan, G.V. Homeowners’ decision-making in a premise plumbing failure-prone area. *J. Am. Water Works Assoc.* 2013, 105, E236–E241. [CrossRef]

74. Ramamurthy, T.; Ghosh, A.; Pazhani, G.P.; Shinoda, S. Current Perspectives on Viable but Non-Culturable (VBNC) Pathogenic Bacteria. *Front. Public Health* 2014, 2, 103. [CrossRef] [PubMed]

75. Alleron, L.; Khemiri, A.; Koubar, M.; Lacombe, C.; Coquet, L.; Cosette, P.; Jouenne, T.; Frere, J. VBNC Legionella pneumophila cells are still able to produce virulence proteins. *Water Res.* 2013, 47, 6606–6617. [CrossRef] [PubMed]

76. Dietersdorfer, E.; Kirschner, A.; Schrammel, B.; Ohradanova-Repic, A.; Stockinger, H.; Sommer, R.; Walochań, J.; Cervero-Aragó, S. Starved viable but non-cultur able (VBNC) Legionella strains can infect and replicate in amoebae and human macrophages. *Water Res.* 2018, 141, 428–438. [CrossRef] [PubMed]

77. Copper Development Association Inc. Copper Facts. Available online: https://www.copper.org/education/c-facts/facts-print.html (accessed on 18 August 2020).

78. Copper Development Association Inc. Antimicrobial Copper Surfaces. Available online: https://www.copper.org/applications/antimicrobial/ (accessed on 18 August 2020).

79. European Chemicals Agency. Information on Biocides. Available online: https://www.echa.europa.eu/web/guest/information-on-chemicals/biocidal-active-substances?p_p_id=dissactivesubstances_WAR_dissactivesubstancesportlet&pp_p_lifecycle=0&pp_p_state=normal&pp_mode=view&pp_p_col_id=column-1&pp_p_col_pos=2&pp_p_col_count=3&dissacti (accessed on 10 November 2020).

80. Samanovic, M.I.; Ding, C.; Thiele, D.J.; Darwin, K.H. Copper in microbial pathogenesis: Meddling with the metal. *Cell Host Microbe* 2012, 11, 106–115. [CrossRef] [PubMed]

81. Boulay, N.; Edwards, M. Copper in the Urban Water Cycle. *Crit. Rev. Environ. Sci. Technol.* 2000, 30, 297–326. [CrossRef]

82. World Health Organization. *Copper in Drinking-Water*; World Health Organization: Geneva, Switzerland, 2016.

83. Lin, Y.S.; Stout, E.J.; Vicror, L.Y. Negative effect of high pH on biocidal efficacy of copper and silver ions in controlling Legionella pneumophila. *Appl. Environ. Microbiol.* 2002, 68, 2711–2715. [CrossRef]

84. Lin, Y.S.; Stout, J.E.; Yu, V.L.; Vidic, R.D. Disinfection of water distribution systems for Legionella. *Semin. Respir. Infect.* 1998, 13, 147–159.

85. Hajialilo, E.; Niyiyati, M.; Solaymani, M.; Rezaeian, M. Pathogenic Free-Living Amoebae Isolated From Contact Lenses of Keratitis Patients. *Iran. J. Parasitol.* 2015, 10, 541–546.

86. Centers for Disease Control and Prevention, National Center for Emerging and Zoonotic Infectious Diseases (NCEZID), Division of Foodborne, Waterborne, and E.D. (DFWED) Parasites—Naegleria fowleri—Primary Amebic Meningoencephalitis (PAM)—Amebic Encephalitis. Available online: https://www.cdc.gov/parasites/naegleria/general.html (accessed on 21 July 2020).

87. Centers for Disease Control and Prevention, National Center for Emerging and Zoonotic Infectious Diseases (NCEZID), Division of Foodborne, Waterborne, and E.D. (DFWED) Parasites—Acanthamoeba—Granulomatous Amebic Encephalitis (GAE). Keratitis. Available online: https://www.cdc.gov/parasites/acanthamoeba/gen_info/acanthamoeba.html (accessed on 10 August 2020).

88. Thomas, V.; Bouchez, T.; Nicolas, V.; Robert, S.; Loret, J.F.; Lévi, Y. Amoebae in domestic water systems: Resistance to disinfection treatments and implication in Legionella persistence. *J. Appl. Microbiol.* 2004, 97, 950–963. [CrossRef] [PubMed]

89. Peleg, A.Y.; Seifert, H.; Paterson, D.L. Acinetobacter baumannii: Emergence of a successful pathogen. *Clin. Microbiol. Rev.* 2008, 21, 538–582. [CrossRef] [PubMed]

90. Cervantes, H.I.; Álvarez, J.A.; Muñoz, J.M.; Arreguin, V.; Mosqueda, J.L.; Macias, A.E. Antimicrobial activity of copper against organisms in aqueous solution: A case for copper-based water pipelines in hospitals? *Am. J. Infect. Control* 2013, 41, e115–e118. [CrossRef] [PubMed]

91. Huang, H.-L.; Shih, H.-Y.; Lee, C.-M.; Yang, T.C.; Lay, J.J.; Lin, Y.E. In vitro efficacy of copper and silver ions in eradicating *Pseudomonas aeruginosa, Stenotrophomonas maltophilia* and *Acinetobacter baumannii*: Implications for on-site disinfection for hospital infection control. *Water Res.* 2008, 42, 73–80. [CrossRef]
92. Shih, H.-Y.Y.; Lin, Y.E. Efficacy of Copper-Silver Ionization in Controlling Biofilm- and Plankton-Associated Waterborne Pathogens. Appl. Environ. Microbiol. 2010, 76, 2032–2035. [CrossRef]
93. Cateau, E.; Delafont, V.; Hechard, Y.; Rodier, M.H. Free-living amoebae: What part do they play in healthcare-associated infections? J. Hosp. Infect. 2014, 87, 131–140. [CrossRef]
94. Cateau, E.; Verdon, J.; Fernandez, B.; Hechard, Y.; Rodier, M.-H. Acanthamoeba sp. promotes the survival and growth of Acinetobacter baumannii. FEMS Microbiol. Lett. 2011, 319, 19–25. [CrossRef]
95. Centers for Disease Control and Prevention, National Center for Emerging and Zoonotic Infectious Diseases (NCEZID), D. of H.Q.P. (DHQP) Nontuberculous Mycobacteria (NTM) Infections. J. Food Prot. 2011, 74, 1321–1329. [CrossRef]
96. Kozajda, A.; Ježak, K.; Kapsa, A. Airborne Staphylococcus aureus in different environments—a review. Environ. Sci. Pollut. Res. Int. 2019, 26, 34741–34753. [CrossRef]
97. Plipat, N.; Spicknall, I.H.; Koopman, J.S.; Eisenberg, J.N. The dynamics of methicillin-resistant Staphylococcus aureus exposure in a hospital model and the potential for environmental intervention. BMC Infect. Dis. 2013, 13, 595. [CrossRef]
98. Landeen, L.K.; Yahya, M.T.; Kutz, S.M.; Gerba, C.P. Microbiological Evaluation of Copper: Silver Disinfection Units for Use in Swimming Pools. Water Sci. Technol. 1989, 21, 267–270. [CrossRef]
99. Yahya, M.T.; Landeen, L.K.; Messina, M.C.; Kutz, S.M.; Schulze, R.; Gerba, C.P. Disinfection of bacteria in water systems by using electrolytically generated coppersonder and reduced levels of free chlorine. Can. J. Microbiol. 1990, 36, 109–116. [CrossRef] [PubMed]
100. Hopkin, M. MRSA “hiding in hospital sinks and vases”. Nature 2006. [CrossRef]
101. Huws, S.A.; Smith, A.W.; Enright, M.C.; Wood, P.J.; Brown, M.R.W. Amoebae promote persistence of epidemic strains of MRS. Environ. Microbiol. 2006, 8, 1130–1133. [CrossRef] [PubMed]
102. Brooke, J.S. Stenotrophomonas maltophilia: An emerging global opportunistic pathogen. Clin. Microbiol. Rev. 2012, 25, 2–41. [CrossRef] [PubMed]
103. Denton, M.; Kerr, K.G. Microbiological and clinical aspects of infection associated with Stenotrophomonas maltophilia. Clin. Microbiol. Rev. 1998, 11, 57–80. [CrossRef]
104. Gomes, I.B.; Simões, L.C.; Simões, M. Influence of surface copper content on Stenotrophomonas maltophilia biofilm control using chlorine and mechanical stress. Biofouling 2020, 36, 1–13. [CrossRef]
105. Thomas, J.M.; Ashbolt, N.J. Do free-living amoebae in treated drinking water systems present an emerging health risk? Environ. Sci. Technol. 2011, 45, 860–869. [CrossRef]
106. Centers for Disease Control and Prevention, National Center for Emerging and Zoonotic Infectious Diseases (NCEZID), D. of H.Q.P. (DHQP) Staphylococcus aureus in Healthcare Settings|HAI|CDC. Available online: https://www.cdc.gov/hai/organisms/staph.html (accessed on 10 August 2020).
107. Johnson, M.M.; Odell, J.A. Nontuberculous mycobacterial pulmonary infections. J. Thorac. Dis. 2014, 6, 210–220. [CrossRef]
108. Kusnetsov, J.; Iivanainen, E.; Elomaa, N.; Martikainen, P.J. Copper and silver ions more effective against Legionella than against mycobacteria in a hospital warm water system. Water Res. 2001, 35, 4217–4225. [CrossRef]
109. Lin, Y.E.; Vidic, R.D.; Stout, J.E.; McCartney, C.A.; Yu, V.L. Inactivation of Mycobacterium avium by copper and silver ions. Water Res. 1998, 32, 1997–2000. [CrossRef]
110. Rhoads, W.J.; Pruden, A.; Edwards, M.A. Interactive Effects of Corrosion, Copper, and Chloramines on Legionella and Mycobacteria in Hot Water Plumbing. Environ. Sci. Technol. 2017, 51, 7065–7075. [CrossRef] [PubMed]
111. Mullis, S.N.; Falkinham, J.O. Adherence and biofilm formation of Mycobacterium avium, Mycobacterium intracellulare and Mycobacterium abscessus to household plumbing materials. J. Appl. Microbiol. 2013, 115, 908–914. [CrossRef] [PubMed]
112. Cirillo, J.D.; Falkow, S.; Tompkins, L.S.; Bermudez, L.E. Interaction of Mycobacterium avium with environmental amoebae enhances virulence. Infect. Immun. 1997, 65, 3759–3767. [CrossRef] [PubMed]
113. Igbinosa, I.H.; Igumbor, E.U.; Aghdasi, F.; Tom, M.; Okoh, A.I. Emerging Aeromonas species infections and their significance in public health. ScientificWorldJournal. 2012, 2012, 625023. [CrossRef]
114. Assant, M.A.; Roy, D.; Montpetit, D. Adhesion of Aeromonas hydrophila to Water Distribution System Pipes after Different Contact Times. J. Food Prot. 1998, 61, 1321–1329. [CrossRef]
115. Rahman, M.; Abd, H.; Romling, U.; Sandstrom, G.; Möllby, R. Aeromonas-Acanthamoeba interaction and early shift to a viable but nonculturable state of Aeromonas by Acanthamoeba. J. Appl. Microbiol. 2008, 104, 1449–1457. [CrossRef]

116. Delafont, V.; Perraud, E.; Brunet, K.; Maisonneuve, E.; Kaaki, S.; Rodier, M.H. Vermamoeba vermiformis in hospital network: A benefit for Aeromonas hydrophila. Parasitol. Res. 2019, 118, 3191–3194. [CrossRef]

117. National Center for Immunization and Respiratory Diseases, D. of B.D. Legionnaires Disease Signs and Symptoms|Legionella|CDC. Available online: https://www.cdc.gov/legionella/about/signs-symptoms.html (accessed on 10 August 2020).

118. National Center for Immunization and Respiratory Diseases, D. of B.D. Legionnaires Disease Cause and Spread|Legionella|CDC. Available online: https://www.cdc.gov/legionella/about/causes-transmission.html (accessed on 10 August 2020).

119. Stout, J.E.; Lin, Y.-S.E.; Goetz, A.M.; Muder, R.R. Controlling Legionella in hospital water systems: Experience with the superheat-and-flush method and copper-silver ionization. Infect. Control Hosp. Epidemiol. 1998, 19, 911–914. [CrossRef] [PubMed]

120. Miettunen, S.; Schwille, R.C.; Farley, A.; Wald, E.R.; Ge, J.H.; States, S.J.; Libert, T.; Wadowsky, R.M. Efficacy of thermal treatment and copper-copper ionization for controlling Legionella pneumophila in high-volume hot water plumbing systems in hospitals. Am. J. Infect. Control 1997, 25, 452–457. [CrossRef]

121. Gião, M.S.; Wilks, S.A.; Keevil, C.W. Influence of copper surfaces on biofilm formation by Legionella pneumophila in potable water. Biometals 2015, 28, 329. [CrossRef] [PubMed]

122. Buse, H.Y.; Lu, J.; Struemwing, I.T.; Ashbolt, N.J. Preferential colonization and release of Legionella pneumophila from mature drinking water biofilms grown on copper versus unplasticized polyvinylchloride coupons. Int. J. Hyg. Environ. Health 2014, 217, 219–225. [CrossRef] [PubMed]

123. Rowbotham, T.J. Preliminary report on the pathogenicity of Legionella pneumophila for freshwater and soil amoebae. J. Clin. Pathol. 1980, 33, 1179–1183. [CrossRef]

124. Grossi, M.; Dey, R.; Ashbolt, N. Searching for Activity Markers that Approximate (VBNC) Legionella pneumophila Infectivity in Amoeba after Ultraviolet (UV) Irradiation. Water 2018, 10, 1219. [CrossRef]

125. Bodey, G.P.; Bolivar, R.; Fainstein, V.; Jadeja, L. Infections Caused by Pseudomonas aeruginosa. Rev. Infect. Dis. 1983, 5, 279–313. [CrossRef] [PubMed]

126. Williams, B.J.; Dehnrostel, J.; Blackwell, T.S. Pseudomonas aeruginosa: Host defence in lung diseases. Respir Med 2010, 15, 1037–1056. [CrossRef] [PubMed]

127. Bédard, E.; Charron, D.; Lalancette, C.; Déziel, E.; Prévost, M. Recovery of Pseudomonas aeruginosa culturability following copper- and chlorine-induced stress. FEMS Microbiol. Lett. 2014, 356, 226–234. [CrossRef]

128. Moritz, M.M.; Flemming, H.C.; Wingender, J. Integration of Pseudomonas aeruginosa and Legionella pneumophila in drinking water biofilms grown on domestic plumbing materials. Int. J. Hyg. Environ. Health 2010, 213, 190–197. [CrossRef]

129. Jeanvoine, A.; Meunier, A.; Puja, H.; Bertrand, X.; Valot, B.; Hocquet, D. Contamination of a hospital plumbing system by persister cells of a copper-tolerant high-risk clone of Pseudomonas aeruginosa. J. Appl. Microbiol. 2010, 109, 1449–1456. [CrossRef] [PubMed]

130. Petignat, C.; Francioli, P.; Nahimana, I.; Wenger, A.; Bille, J.; Schaller, M.-D.; Revelly, J.-P.; Zanetti, G.; Blanc, D.S. Exogenous Sources of Pseudomonas aeruginosa in Intensive Care Unit Patients Implementation of Infection Control Measures and Follow-Up With Molecular Typing. Infect. Control Hosp. Epidemiol. 2006, 27, 953–957. [CrossRef]

131. Teitzel, G.M.; Parsek, M.R. Heavy Metal Resistance of Biofilm and Planktonic Pseudomonas aeruginosa. Appl. Environ. Microbiol. 2003, 69, 2313. [CrossRef]

132. Dvidjosiskojo, Z.; Richard, J.; Moritz, M.M.; Dopp, E.; Flemming, H.-C.; Wingender, J. Influence of copper ions on the viability and cytotoxicity of Pseudomonas aeruginosa under conditions relevant to drinking water environments. Int. J. Hyg. Environ. Health 2011, 214, 485–492. [CrossRef] [PubMed]

133. Dopp, E.; Richard, J.; Dvidjosiskojo, Z.; Simon, A.; Wingender, J. Influence of the copper-induced viable but non-culturable state on the toxicity of Pseudomonas aeruginosa towards human bronchial epithelial cells in vitro. Int. J. Hyg. Environ. Health 2017, 220, 1363–1369. [CrossRef] [PubMed]
134. Pukatzki, S.; Kessin, R.H.; Mekalanos, J.J. The human pathogen Pseudomonas aeruginosa utilizes conserved virulence pathways to infect the social amoeba Dictyostelium discoideum. *Proc. Natl. Acad. Sci. USA* 2002, 99, 3159. [CrossRef] [PubMed]

135. US EPA Office of Water. *Technologies for Legionella Control in Premise Plumbing Systems*; US Environmental Protection Agency: Washington, DC, USA, 2016.

136. Hans, M.; Erbe, A.; Mathews, S.; Chen, Y.; Soli, M.; Mücklich, F. Role of Copper Oxides in Contact Killing of Bacteria. *Langmuir* 2013, 29, 16160–16166. [CrossRef] [PubMed]

137. Cachafeiro, S.P.; Naveira, I.M.; García, I.G. Is copper-silver ionisation safe and effective in controlling legionella? *J. Hosp. Infect.* 2007, 67, 209–216. [CrossRef]

138. Kim, B.R.; Anderson, J.E.; Mueller, S.A.; Gaines, W.A.; Kendall, A.M. Literature review—Efficacy of various disinfectants against Legionella in water systems. *Water Res.* 2002, 36, 4433–4444. [CrossRef]

139. Rakshit, A.; Khatua, K.; Shanbhag, V.; Comba, P.; Datta, A. Cu2+ selective chelators relieve copper-induced oxidative stress: In vivo. *Chem. Sci.* 2018, 9, 7916–7930. [CrossRef]

140. Van der Kooij, D.; Veemendaal, H.R.; Scheffer, W.J.H. Biofilm formation and multiplication of Legionella in a model warm water system with pipes of copper, stainless steel and cross-linked polyethylene. *Water Res.* 2005, 39, 2789–2798. [CrossRef]

141. Dodrill, D.M.; Edwards, M. Corrosion control on the basis of utility experience. *J. Am. Water Works Assoc.* 1995, 87, 74–85. [CrossRef]

142. Kimbrough, D.E. Brass corrosion as a source of lead and copper in traditional and all-plastic distribution systems. *J. Am. Water Works Assoc.* 2007, 99, 70–76. [CrossRef]

143. June, S.G.; Dżiwułski, D.M. Copper and Silver Biocidal Mechanisms, Resistance Strategies, and Efficacy for Legionella Control. *Journal-American Water Work. Assoc.* 2018, 110, E13–E35. [CrossRef]

144. Assaidi, A.; Ellouali, M.; Latrache, H.; Mabrouki, M.; Hamadi, F.; Timinouni, M.; Zahir, H.; El Mdaghri, N.; Barguiguia, A.; Miji, E.M. Effect of temperature and plumbing materials on biofilm formation by Legionella pneumophila serogroup 1 and 2. *J. Adhes. Sci. Technol.* 2018, 32, 1471–1484. [CrossRef]

145. Ashbolt, N.J. Environmental (Saprozoic) Pathogens of Engineered Water Systems: Understanding Their Ecology for Risk Assessment and Management. *Pathog.* 2015, 4, 390–405. [CrossRef] [PubMed]

146. Morvay, A.A.; Decun, M.; Scurtu, M.; Sala, C.; Morar, A.; Sarandan, M. Biofilm formation on materials commonly used in household drinking water systems. *Water Sci. Technol. Water Supply* 2011, 11, 252–257. [CrossRef]

147. Dai, D.; Proctor, C.R.; Williams, K.; Edwards, M.A.; Pruden, A. Mediation of effects of biofiltration on bacterial regrowth, Legionella pneumophila, and the microbial community structure under hot water plumbing conditions. *Environ. Sci. Water Res. Technol.* 2018, 4, 183–194. [CrossRef]

148. Soothill, J.S. Carbapenemase-bearing Klebsiella spp. in sink drains: Investigation into the potential advantage of copper pipes. *J. Hosp. Infect.* 2016, 93, 152–154. [CrossRef]

149. Norton, C.D.; LeChevallier, M.W.; Falkinham, J.O. Survival of Mycobacterium avium in a model distribution system. *Water Res.* 2004, 38, 1457–1466. [CrossRef]

150. Leoni, E.; De Luca, G.; Legnani, P.P.; Sacchetti, R.; Stampi, S.; Zanetti, F. Legionella waterline colonization: Detection of Legionella species in domestic, hotel and hospital hot water systems. *J. Appl. Microbiol.* 2005, 98, 373–379. [CrossRef]

151. Marrie, T.; Green, P.; Burbridge, S.; Bezanson, G.; Neale, S.; Hoffman, P.S.; Haldane, D. Legionellaceae in the potable water of Nova Scotia hospitals and Halifax residences. *Epidemiol. Infect.* 1994, 112, 143–150. [CrossRef]

152. Borella, P.; Montagna, M.T.; Stampi, S.; Stancanelli, G.; Romano-Spica, V.; Triassi, M.; Marchesi, I.; Bargellini, A.; Tato, D.; Napoli, C. Legionella contamination in hot water of Italian hotels. *Appl. Environ. Microbiol.* 2005, 71, 5805–5813. [CrossRef]

153. Mathys, W.; Holmman, C.P.; Junge-Mathys, E. Efficacy of Copper-Silver Ionization in Controlling Legionella in a Hospital Hot Water Distribution System: A German Experience. In *Legionella*; Marre, R., Kwaik, Y.A., Bartlett, C., Cianciotto, N.P., Fields, B.S., Frosch, M., Hacker, J., Luck, P.C., Eds.; ASM Press: Washington, DC, USA, 2002; pp. 419–424. ISBN 1-55581-230-9.

154. Bédard, E.; Paranjape, K.; Lalancette, C.; Villion, M.; Quach, C.; Laferrière, C.; Faucher, S.P.; Prévost, M. Legionella pneumophila levels and sequence-type distribution in hospital hot water samples from faucets to connecting pipes. *Water Res.* 2019, 156, 277–286. [CrossRef] [PubMed]
155. Bédard, E.; Trigui, H.; Liang, J.; Doberva, M.; Paranjape, K.; Lalancette, C.; Faucher, S.P.; Prévost, M. Local adaptation of Legionella pneumophila within a hospital hot water system increases tolerance to copper. *bioRxiv* 2020. [CrossRef]

156. Lytle, D.A.; Liggett, J. Impact of water quality on chlorine demand of corroding copper. *Water Res.* 2016, 92, 11–21. [CrossRef] [PubMed]

157. Edwards, M.; Powers, K.; Hidmi, L.; Schock, M.R. *The role of Pipe Ageing in Copper Corrosion by-Product Release*; IWA Publishing: London, UK, 2001; Volume 1, pp. 25–32.

158. Zevenhuizen, L.P.T.M.; Dolfing, J.; Eshuis, E.J.; Scholten-Koerselman, I.J. Inhibitory effects of copper on bacteria related to the free ion concentration. *Microb. Ecol.* 1979, 5, 139–146. [CrossRef]

159. Garvey, J.E.; Owen, H.A.; Winner, R.W. Toxicity of copper to the green alga, Chlamydomonas reinhardtii. *J. Electrochem. Soc.* 1991, 19, 89–96. [CrossRef]

160. Meador, J.P. The interaction of pH, dissolved organic carbon, and total copper in the determination of ionic copper and toxicity. *Aquat. Toxicol.* 1991, 19, 13–31. [CrossRef]

161. Ebrahimpour, M.; Alipour, H.; Rakshah, S. Influence of water hardness on acute toxicity of copper and zinc on fish. *Toxicol. Ind. Health* 2010, 26, 361–365. [CrossRef]

162. Pourkhabbaz, A.; Kasmani, M.E.; Kiyani, V.; Hosynzadeh, M.H. Effects of water hardness and Cu and Zn on LC50 in Gambusia holbrooki. *Chem. Speciat. Bioavailab.* 2011, 23, 224–228. [CrossRef]

163. Riethmuller, N.; Markich, S.; Parry, D.; van Dam, R. *The Effect of True Water Hardness and Alkalinity on the Toxicity of Cu and U to Two Tropical Australian Freshwater Organisms*; Supervising Scientist, Environment Australia: Darwin NT, Australia, 2000.

164. Yang, S.; Amy, P.; Edwards, M.; Rhoads, W. Natural Organic Matter, Orthophosphate, pH, and Growth Phase Can Limit Copper Antimicrobial Efficacy for Legionella in Drinking Water. *Environ. Sci. Technol.* 2020. Submitted.

165. Stauber, J.L.; Florence, T.M. Mechanism of toxicity of ionic copper and copper complexes to algae. *Mar. Biol.* 1987, 94, 511–519. [CrossRef]

166. Ives, D.J.G.; Rawson, A.E. Copper Corrosion. *J. Electrochem. Soc.* 1962, 109, 447. [CrossRef]

167. Galai, S.; Touhami, Y.; Marzouki, M.N. Response surface methodology applied to laccases activities exhibited by Stenotrophomonas maltophilia AAP56 in different growth conditions. *BioResources* 2012, 7, 706–726.

168. Wingender, J.; Flemming, H.C. Biofilms in drinking water and their role as reservoir for pathogens. *Int. J. Hyg. Environ. Health* 2011, 214, 417–423. [CrossRef] [PubMed]

169. Van der Kooij, D.; van der Wielen, P.W.J.J.; Rosso, D.; Shaw, A.; Borchardt, D.; Ibsch, R.; Apgar, D.; Witherspoon, J.; Di Toro, D.M.; Paquin, P.R. *Microbial Growth in Drinking Water Supplies*; IWA Publishing: London, UK, 2013; ISBN 1780400403.

170. Hallam, N.B.; West, J.R.; Forster, C.F.; Simms, J. The potential for biofilm growth in water distribution systems. *Water Res.* 2001, 35, 4063–4071. [CrossRef]

171. Tsuchida, D.; Kajihara, Y.; Shimizu, N.; Hamamura, K.; Nagase, M. Hydrogen sulfide production by sulfate-reducing bacteria utilizing additives eluted from plastic resins. *Waste Manag. Res.* 2011, 29, 594–601. [CrossRef]

172. Kelley, K.M. *The Impact of Cross-linked Polyethylene (PEX) Pipe on Drinking Water Chemical and Odor Quality*; University of South Alabama: Mobile, AL, USA, 2014.

173. Corfitzen, C.B. *Afgivelse af Organisk Stof Fra Polymere Materialer—Mikrobiel Vækst [Release of Organic Matter from Polymeric Materials—Microbial Growth]*; The Danish Environmental Protection Agency: Odense, Denmark, 2002.

174. Kowalska, B.; Kowalski, D.; Rozej, A. Organic compounds migrating from plastic pipes into water. *J. Water Supply Res. Technol. AQUA* 2011, 60, 137–146. [CrossRef]

175. Pedersen, K. Biofilm development on stainless steel and pvc surfaces in drinking water. *Water Res.* 1990, 24, 239–243. [CrossRef]

176. Thomson, R.M.; Carter, R.; Tolson, C.; Coulter, C.; Huygens, F.; Hargreaves, M. Factors associated with the isolation of Nontuberculous mycobacteria (NTM) from a large municipal water system in Brisbane, Australia. *BMC Microbiol.* 2013, 13, 89. [CrossRef]
178. Learbuch, K.L.G.; Lut, M.C.; Liu, G.; Smidt, H.; van der Wielen, P.W.J. Legionella growth potential of drinking water produced by a reverse osmosis pilot plant. *Water Res.* **2019**, *157*, 55–63. [CrossRef]

179. Buse, H.Y.; Morris, B.; Struewing, I.T.; Szabo, J.G. Chlorine and monochloramine disinfection of Legionella pneumophila colonizing copper and PVC drinking water biofilms. *Appl. Environ. Microbiol.* **2019**. [CrossRef] [PubMed]

180. Zhang, Y.; Griffin, A.; Edwards, M. Effect of nitrification on corrosion of galvanized iron, copper, and concrete. *J. Am. Water Works Assoc.* **2010**, *102*, 83–93. [CrossRef]

181. Sarin, P.; Snoeyink, V.; Bebee, J.; Jim, K.K.; Beckett, M.A.; Kriven, W.M.; Clement, J.A. Iron release from corroded iron pipes in drinking water distribution systems: Effect of dissolved oxygen. *Water Res.* **2004**, *38*, 1259–1269. [CrossRef]

182. WHO. *Legionella and the Prevention of Legionellosis*; World Health Organization: Geneva, Switzerland, 2007.

183. Faraldo-Gómez, J.D.; Sansom, M.S.P. Acquisition of siderophores in gram-negative bacteria. *Nat. Rev. Mol. Cell Biol.* **2003**, *4*, 105–116. [CrossRef]

184. States, S.J.; Conley, L.F.; Ceraso, M.; Stephenson, T.E.; Wolford, R.S.; Wadowsky, R.M.; McNamara, A.M.; Yee, R.B. Effects of metals on Legionella pneumophila growth in drinking water plumbing systems. *Appl. Environ. Microbiol.* **1985**, *50*, 1149–1154. [CrossRef] [PubMed]

185. Rakić, A.; Perić, J.; Foglar, L. Influence of temperature, chlorine residual and heavy metals on the presence of *Legionella pneumophila* in hot water distribution systems. *Ann. Agric. Environ. Med.* **2012**, *19*, 431–436. [PubMed]

186. van der Lugt, W.; Euser, S.M.; Bruin, J.P.; Den Boer, J.W.; Walker, J.T.; Crespi, S. Growth of Legionella anisa in a model drinking water system to evaluate different shower outlets and the impact of cast iron rust. *Int. J. Hggy. Environ. Health* **2017**, *220*, 1295–1308. [CrossRef] [PubMed]

187. Martin, R.; Strom, O.; Pruden, A.; Edwards, M. Copper Pipe, Stagnant Conditions, Corrosion Control, and Trace Disinfectant Residual Enhanced Reduction of Legionella pneumophila during Simulations of the Flint Water Crisis. *Pathogens* **2020**, in press. [CrossRef]

188. Wang, H.; Masters, S.; Hong, Y.; Stallings, J.; Falkinham, J.O.; Edwards, M.A.; Pruden, A. Effect of Disinfectant, Water Age, and Pipe Material on Occurrence and Persistence of Legionella, mycobacteria, Pseudomonas aeruginosa, and Two Amoebas. *Environ. Sci. Technol.* **2012**, *46*, 11566–11574. [CrossRef]

189. Oyem, H.H.; Oyem, I.M.; Usese, A.I. Iron, manganese, cadmium, chromium, zinc and arsenic groundwater contents of Agbor and Owa communities of Nigeria. *Springerplus* **2015**, *4*, 104. [CrossRef]

190. World Health Organization. *Zinc in Drinking Water: Background Document for Development of Guidelines for Drinking-Water Quality. Health Criteria and Other Supporting Information*; World Health Organization: Geneva, Switzerland, 2003; Volume 1.

191. World Health Organization Water Sanitation. *Guidelines for Drinking-Water Quality: Incorporating First Addendum. Vol. 1, Recommendations*; World Health Organization: Geneva, Switzerland, 2006.

192. Salehi, M.; Odimayomi, T.; Ra, K.; Ley, C.; Julien, R.; Nejadhashemi, A.P.; Hernandez-Suarez, J.S.; Mitchell, J.; Shah, A.D.; Whelan, A. An investigation of spatial and temporal drinking water quality variation in green residential plumbing. *Build. Environ.* **2020**, *169*, 106566. [CrossRef]

193. Sharrett, A.R.; Carter, A.P.; Orheimt, R.M.; Feinleib, M. Daily intake of lead, cadmium, copper, and zinc from drinking water: The seattle study of trace metal exposure. *Environ. Res.* **1982**, *28*, 456–475. [CrossRef]

194. Howard, C.D. Zinc Contamination in Drinking Water. *J. Am. Water Works Assoc.* **1923**, *10*, 411–414. [CrossRef]

195. Silver, S.; Lusk, J.E. Bacterial Magnesium, Manganese, and Zinc Transport. In *Ion Transport in Prokaryotes*; Rosen, B.P., Silver, S., Eds.; Academic Press: Cambridge, MA, USA, 1987; pp. 165–180. ISBN 978-0-12-596935-2.

196. Choudhury, R.; Srivastava, S. Zinc resistance mechanisms in bacteria. *Curr. Sci.* **2001**, *81*, 768–775.

197. Ma, L.; Terwilliger, A.; Maresco, A.W. Iron and zinc exploitation during bacterial pathogenesis. *Metallomics* **2015**, *7*, 1541–1554. [CrossRef]

198. Li, Y.; Sharma, M.R.; Koripella, R.K.; Yang, Y.; Kaushal, P.S.; Lin, Q.; Wade, J.T.; Gray, T.A.; Derbyshire, K.M.; Agrawal, R.K.; et al. Zinc depletion induces ribosome hibernation in mycobacteria. *Proc. Natl. Acad. Sci. USA* **2018**, *115*, 8191. [CrossRef]

199. Suryawati, B. Zinc homeostasis mechanism and its role in bacterial virulence capacity. *AIP Conf. Proc.* **2018**, *2021*, 070021. [CrossRef]

200. Nies, D.H. Resistance to cadmium, cobalt, zinc, and nickel in microbes. *Plasmid* **1992**, *27*, 17–28. [CrossRef]
201. Reeves, M.W.; Pine, L.; Hutner, S.H.; George, J.R.; Harrell, W.K. Metal requirements of Legionella pneumophila. *J. Clin. Microbiol.* 1981, 13, 688–695. [CrossRef]

202. Kirschner, R.A.; Parker, B.C.; Falkinham, J.O. Epidemiology of Infection by Nontuberculous Mycobacteria: Mycobacterium avium, Mycobacterium intracellulare, and Mycobacterium scrofulaceum in Acid, Brown-Water Swamps of the Southeastern United States and Their Association with Environmental Variables. *Am. Rev. Respir. Dis.* 1992, 145, 271–275. [CrossRef]

203. Norberg, A.B.; Molin, N. Toxicity of cadmium, cobalt, uranium and zinc to Zoogloea ramigera. *Water Res.* 1983, 17, 1333–1336. [CrossRef]

204. Nweke, C.; Chibuogwu, O.; Alisi, C. Response of planktomic bacteria of New Calabar River to zinc stress. *Afr. J. Biotechnol.* 2006, 5, 653–658.

205. Nweke, C.; Alisi, C.; Chibuogwu, O.; Nwanyanwu, C. Toxicity of Zinc to heterotrophic bacteria from a tropical river sediment. *Appl. Ecol. Environ. Res.* 2007, 5, 123–132. [CrossRef]

206. Babich, H.; Stotzky, G. Toxicity of zinc to fungi, bacteria, and coliphages: Influence of chloride ions. *Appl. Environ. Microbiol.* 1978, 36, 906. [CrossRef]

207. United States Environmental Protection Agency. *Secondary Drinking Water Standards: Guidance for Nuisance Chemicals;* United States Environmental Protection Agency: Washington, DC, USA, 2015.

208. Standardization Administration of China Guobiao 5749-2006. Available online: https://www.chinesestandard.net/PDF.aspx?gB5749-2006 (accessed on 10 November 2020).

209. Hwang, M.G.; Katayama, H.; Ohgaki, S. Effect of Intracellular Resuscitation of Legionella pneumophila in Acanthamoeba polyphage Cells on the Antimicrobial Properties of Silver and Copper. *Environ. Sci. Technol.* 2006, 40, 7434–7439. [CrossRef]

210. Barker, J.; Brown, M.R.; Collier, P.J.; Farrell, I.; Gilbert, P. Relationship between Legionella pneumophila and Acanthamoeba polyphaga: Physiological status and susceptibility to chemical inactivation. *Appl. Environ. Microbiol.* 1992, 58, 2420–2425. [CrossRef]

211. García, M.T.; Jones, S.; Pelaz, C.; Millar, R.D.; Abu Kwaik, Y. Acanthamoeba polyphaga resuscitates viable non-culturab Legionella pneumophila after disinfection. *Environ. Microbiol.* 2007, 9, 1267–1277. [CrossRef]

212. Adékambi, T; Ben Salah, S; Khilif, M; Raoul, D; Drancourt, M. Survival of Environmental Mycobacteria in Acanthamoeba polyphaga. *Appl. Environ. Microbiol.* 2006, 72, 5974. [CrossRef]

213. Cervero-Aragó, S.; Rodriguez-Martínez, S.; Puertas-Bennasar, A.; Araujo, R.M. Effect of Common Drinking Water Disinfectants, Chlorine and Heat, on Free Legionella and Amoebae-Associated Legionella. *PLoS ONE* 2015, 10, e0134726. [CrossRef]

214. Marciano-Cabral, F; Jamerson, M.; Kaneshiro, E.S. Free-living amoebae, Legionella and Mycobacterium in tap water supplied by a municipal drinking water utility in the USA. *J. Water Health* 2009, 8, 71–82. [CrossRef]

215. Inkinen, J; Jayaprakash, B; Ahonen, M; Pitkänen, T; Mäkinen, R; Pursiainen, A; Santo Domingo, J.W.; Salonen, H; Elk, M; Keinänen-Toivola, M.M. Bacterial community changes in copper and PEX drinking water pipeline biofilms under extra disinfection and magnetic water treatment. *J. Appl. Microbiol.* 2018, 124, 611–624. [CrossRef][PubMed]

216. Vikesland, P.J.; Valentine, R.L. Reaction Pathways Involved in the Reduction of Monochloramine by Ferrous *Environ. Sci. Technol.* 2000, 34, 83–90. [CrossRef]

217. Zhang, Z.; Stout, J.E.; Yu, V.L.; Vidic, R. Effect of pipe corrosion scales on chlorine dioxide consumption in drinking water distribution systems. *Water Res.* 2008, 42, 129–136. [CrossRef][PubMed]

218. Nguyen, C.K.; Powers, K.A.; Raetz, M.A.; Parks, J.L.; Edwards, M.A. Rapid free chlorine decay in the presence of Cu(OH)2: Chemistry and practical implications. *Water Res.* 2011, 45, 5302–5312. [CrossRef][PubMed]

219. Fu, J.; Qu, J.; Liu, R.; Zhao, X.; Qiang, Z. The influence of Cu(I) on the decay of monochloramine. *Chemosphere* 2009, 74, 181–186. [CrossRef][PubMed]

220. Edwards, M.A.; Parks, J.; Griffin, A.; Raetz, M.A.; Martin, A.K.; Scardina, P.; Elfland, C. *Lead and Copper Corrosion Control in New Construction;* The Water Research Foundation: Denver, CO, USA, 2011; ISBN 978-1-60573-137-7.

221. Brazeau, R.H.; Edwards, M.A. Role of Hot Water System Design on Factors Influential to Pathogen Regrowth: Temperature, Chlorine Residual, Hydrogen Evolution, and Sediment. *Environ. Eng. Sci.* 2013, 30, 617–627. [CrossRef]

222. Ishizaki, A.; Tanaka, K.; Taga, N. Microbial production of poly-D-3-hydroxybutyrate from CO2x. *Appl. Microbiol. Biotechnol.* 2001, 57, 6–12. [CrossRef]
223. Edwards, M.; Pruden, A.; Falkinham III, J.O.; Brazeau, R.; Williams, K.; Wang, H.; Martin, A.; Rhoads, W. Relationship between Biodegradable Organic Matter and Pathogen Concentrations in Premise Plumbing. Available online: https://www.researchgate.net/publication/282650492_Relationship_Between_Biodegradable_Organic_Matter_and_Pathogen_Concentrations_in_Premise_Plumbing (accessed on 7 October 2020).

224. Dai, D.; Rhoads, W.J.; Edwards, M.A.; Pruden, A. Shotgun Metagenomics Reveals Taxonomic and Functional Shifts in Hot Water Microbiome Due to Temperature Setting and Stagnation. Front. Microbiol. 2018, 9, 2695. [CrossRef]

225. González-Cabaleiro, R.; Curtis, T.P.; Oﬁçu, I.D. Bioenergetics analysis of ammonia-oxidizing bacteria and the estimation of their maximum growth yield. Water Res. 2019, 154, 238–245. [CrossRef]

226. Neubauer, S.C.; Emerson, D.; Megonigal, J.P. Life at the energetic edge: Kinetics of circumneutral iron oxidation by lithotrophic iron-oxidizing bacteria isolated from the wetland-plant rhizosphere. Appl. Environ. Microbiol. 2002, 68, 3988–3995. [CrossRef]

227. Martin, A.K.; Edwards, M.A.; Pruden, A.J.; Falkinham, J.O. Organic Carbon Generation Mechanisms in Main and Premise Distribution Systems; Virginia Tech: Blacksburg, VA, USA, 2012.

228. Clark, B.; Clair, J.S.; Edwards, M. Copper Deposition Corrosion Elevates Lead Release to Potable Water. J. Am. Water Works Assoc. 2015, 107, E627–E637. [CrossRef]

229. Brazeau, R.H.; Edwards, M.A.; Falkinham, J.O.; Pearce, A.R.; Pruden, A.J. Sustainability of Residential Hot Water Infrastructure: Public Health, Environmental Impacts, and Consumer Drivers; Virginia Tech: Blacksburg, VA, USA, 2012.

230. Fox, P.; Abbasszadegan, M. Impact of Scale Formation on Biofilm Growth in Premise Plumbing. Available online: https://www.semanticscholar.org/paper/Impact-of-Scale-Formation-on-Biofilm-Growth-in-Fox-Abbasszadegan/b615f8d2516a1bac2ac11854019a654e6fc2 (accessed on 31 August 2020).

231. Sarin, P.; Snoeyink, V.L.; Lytle, D.A.; Kriven, W.M. Iron Corrosion Scales: Model for Scale Growth, Iron Release, and Colored Water Formation. J. Environ. Eng. 2004, 130, 364–373. [CrossRef]

232. Yang, F.; Shi, B.; Gu, J.; Wang, D.; Yang, M. Morphological and physicochemical characteristics of iron corrosion scales formed under different water source histories in a drinking water distribution system. Water Res. 2012, 46, 5423–5433. [CrossRef] [PubMed]

233. Tuvoinen, O.H.; Button, K.S.; Vuorinen, A.; Carlson, L.; Mair, D.M.; Yut, L.A. Bacterial, chemical, and mineralogical characteristics of tubercles in distribution pipelines. J. Am. Water Works Assoc. 1980, 72, 626–635. [CrossRef]

234. Lee, W.; Lewandowski, Z.; Nielsen, P.H.; Allan Hamilton, W. Role of sulfate-reducing bacteria in corrosion of mild steel: A review. Biofouling 1995, 8, 165–194. [CrossRef]

235. Nguyen, C.K.; Stone, K.R.; Dudi, A.; Edwards, M.A. Corrosive Microenvironments at Lead Solder Surfaces Arising from Galvanic Corrosion with Copper Pipe. Environ. Sci. Technol. 2010, 44, 7076–7081. [CrossRef]

236. Hasit, Y.J.; Anderson, J.L.; Anthony, J.; Parolari, T.; Rockaway, T.D.; Frenc, M.L. Distribution Water Quality Issues Related to New Development or Low Usage; Awwa Research Foundation: Denver, CO, USA, 2006.

237. McNeill, L.S.; Edwards, M. Iron pipe corrosion in distribution systems. J. Am. Water Works Assoc. 2001, 93, 88–100. [CrossRef]

238. Nawrocki, J.; Raczyk-Stanislawiak, U.; Świętki, J.; Olejnik, A.; Srok, M.J. Corrosion in a distribution system: Steady water and its composition. Water Res. 2010, 44, 1863–1872. [CrossRef]

239. Merkel, T.H.; Groß, H.J.; Werner, W.; Dahlke, T.; Reicherter, S.; Beuchle, G.; Eberle, S.H. Copper corrosion by-product release in long-term stagnation experiments. Water Res. 2002, 36, 1547–1555. [CrossRef]

240. Lytle, D.A.; Schock, M.R. Impact of stagnation time on metal dissolution from plumbing materials in drinking water. J. Water Supply Res. Technol. AQUA 2000, 49, 243–257. [CrossRef]

241. Zhang, L.; Liu, S. Investigation of organic compounds migration from polymeric pipes into drinking water under long retention times. Procedia Eng. 2014, 70, 1753–1761. [CrossRef]

242. Richard, R.; Hamilton, K.A.; Westerhoff, P.; Boyer, T.H. Tracking copper, chlorine, and occupancy in a new, multi-story, institutional green building. Environ. Sci. Water Res. Technol. 2020, 6, 1672–1680. [CrossRef]

243. Haig, S.J.; Kotlarz, N.; Lipuma, J.J.; Raskin, L. A high-throughput approach for identification of nontuberculous mycobacteria in drinking water reveals relationship between water age and Mycobacterium avium. MBio 2018, 9. [CrossRef] [PubMed]
244. Ley, C.; Proctor, C.; Singh, G.; Ra, K.; Nob, Y.; Odimayomi, T.; Salehi, M.; Julien, R.; Mitchell, J.; Nejadhashemi, A.P.; et al. Drinking water microbiology in a water-efficient building: Stagnation, seasonality, and physiochemical effects on opportunistic pathogen and total bacteria proliferation. *Environ. Sci. Water Res. Technol.* 2020. [CrossRef]

245. Hozalski, R.M.; Lapara, T.M.; Zhao, X.; Kim, T.; Waak, M.B.; Burch, T.; Mccarty, M. Flushing of stagnant premise water systems after the COVID-19 shutdown can reduce infection risk by Legionella and Mycobacterium spp. *medRxiv* 2020. [CrossRef]

246. Sánchez-Busó, L.; Olmos, M.P.; Camaró, M.L.; Adrián, F.; Calafat, J.M.; González-Candelas, F. Phylogenetic analysis of environmental Legionella pneumophila isolates from an endemic area (Alcoy, Spain). *Infect. Genet. Evol.* 2015, 30, 45–54. [CrossRef] [PubMed]

247. Pryor, M.; Springthorpe, S.; Riffard, S.; Brooks, T.; Huo, Y.; Davis, G.; Sattar, S.A. Investigation of opportunistic pathogens in municipal drinking water under different supply and treatment regimes. *Water Sci. Technol.* 2004, 50, 83–90. [CrossRef]

248. Guidance for Reopening Buildings After Prolonged Shutdown or Reduced Operation|CDC. Available online: https://www.cdc.gov/coronavirus/2019-ncov/php/building-water-system.html (accessed on 26 October 2020).

249. Proctor, C.; Rhoads, W.; Keane, T.; Salehi, M.; Hamilton, K.; Pieper, K.J.; Cwiertny, D.M.; Prévost, M.; Whelton, A. Considerations for Large Building Water Quality after Extended Stagnation. *AWWA Water Sci.* 2020, 2, e1186. [CrossRef]

250. United States Environmental Protection Agency. Information on Maintaining or Restoring Water Quality in Buildings with Low or No Use. Available online: https://www.epa.gov/coronavirus/information-maintaining-or-restoring-water-quality-buildings-low-or-no-use (accessed on 26 October 2020).

251. Rhoads, W.J.; Prévost, M.; Pieper, K.J.; Keane, T.; Whelton, A.J.; Rölli, F.; Proctor, C.R.; Grimard-Conea, M. Responding to Water Stagnation in Buildings with Reduced or No Water Use; American Water Works Association: Denver, CO, USA, 2020.

252. 2021 Uniform Plumbing Code; Twenty-Nin.; International Association of Plumbing and Mechanical Officials: Ontario, CA, USA, 2020.

253. 2018 International Plumbing Code; International Code Council, Inc.: Country Club Hills, IL, USA, 2017; ISBN 978-1-60983-745-7.

254. 2015 International Energy Conservation Code; International Code Council, Inc: Country Club Hills, IL, USA, 2014; ISBN 978-1-60983-486-9.

255. Ally, M.R. Water and Energy Savings using Demand Hot Water Recirculating Systems in Residential Homes: A Case Study of Five Homes in Palo Alto, California. Available online: http://www.osti.gov/servlets/purl/885864-jNK2Il (accessed on 26 August 2020).

256. Rhoads, W.J.; Ji, P.; Pruden, A.; Edwards, M.A. Water heater temperature set point and water use patterns influence Legionella pneumophila and associated microorganisms at the tap. *Microbiome* 2015, 3, 1–13. [CrossRef]

257. Flannery, B.; Gelling, L.B.; Vugia, D.J.; Weintraub, J.M.; Salerno, J.J.; Conroy, M.J.; Stevens, V.A.; Rose, C.E.; Moore, M.R.; Fields, B.S.; et al. Reducing Legionella colonization of water systems with monochloramine. *Emerg. Infect. Dis.* 2006, 12, 588. [CrossRef]

258. Oh, S.J.; Cook, D.C.; Townsend, H.E. Characterization of Iron Oxides Commonly Formed as Corrosion Products on Steel. *Hyperfine Interact.* 1998, 112, 59–66. [CrossRef]

259. Stout, J.E.; Yu, V.L.; Best, M.G. Ecology of Legionella pneumophila within water distribution systems. *Appl. Environ. Microbiol.* 1985, 49, 221–228. [CrossRef] [PubMed]

260. Zhu, Z.; Wu, C.; Zhong, D.; Yuan, Y.; Shan, L.; Zhang, J. Effects of pipe materials on chlorine-resistant biofilm formation under long-term high chlorine level. *Appl. Biochem. Biotechnol.* 2014, 173, 1564–1578. [CrossRef] [PubMed]

261. Sancy, M.; Gourbeyre, Y.; Sutter, E.M.M.; Tribollet, B. Mechanism of corrosion of cast iron covered by aged corrosion products: Application of electrochemical impedance spectrometry. *Corros. Sci.* 2010, 52, 1222–1227. [CrossRef]

262. Salehi, M.; Abouali, M.; Wang, M.; Zhou, Z.; Nejadhashemi, A.P.; Mitchell, J.; Caskey, S.; Whelton, A.J. Case study: Fixture water use and drinking water quality in a new residential green building. *Chemosphere* 2018. [CrossRef]
263. Lund, V.; Anderson-Glenna, M.; Skjevrak, I.; Steffensen, I.-L. Long-term study of migration of volatile organic compounds from cross-linked polyethylene (PEX) pipes and effects on drinking water quality. *J. Water Health* 2011, 9, 483–497. [CrossRef] [PubMed]

264. Tanemura, D.; Yamada, K.; Nishimura, H.; Igawa, K.; Higuchi, Y. Investigation of degradation mechanism by copper catalytic activity and mechanical property of polyethylene pipes for hot water supply. *Annu. Tech. Conf. ANTEC Conf. Proc.* 2014, 3, 2022–2026.

265. Wright, D. Failure of Plastics and Rubber Products: Causes, Effects and Case Studies Involving Degradation; Rapra Technology Limited: Shawbury, UK, 2001; ISBN 185957517X.

266. Plastic Pipe Institute. *Proper Integration of Copper Tubing and Components with PP-R Piping Materials for Plumbing Applications*; The Plastic Pipe Institute: Irving, TX, USA, 2018.

267. Bulletin, A.T. *Intermixing Copper Tube and Aquatherm*; Aquatherm GmbH: Attendom, Germany, 2012.

268. Huang, X.; Pieper, K.J.; Cooper, H.K.; Diaz-Amaya, S.; Zemlyanov, D.Y.; Whelton, A.J. Corrosion of upstream metal plumbing components impact downstream PEX pipe surface deposits and degradation. *Chemosphere* 2019, 236, 124329. [CrossRef] [PubMed]

269. Lewis, J.R. The Catalytic Decomposition of Sodium Hypochlorite Solutions.II. Iron Oxide as Promoter in the Copper Oxide Catalyst of Sodium Hypochlorite. *J. Phys. Chem.* 1928, 32, 1808–1819. [CrossRef]

270. Cruse, H. Dissolved-Copper Effect on Iron Pipe. *J. Am. Water Works Assoc.* 1971, 63, 79–81. [CrossRef]

271. Cartier, C.; Nour, S.; Richer, B.; Deshommes, E.; Prévost, M. Impact of water treatment on the contribution of faucets to dissolved and particulate lead release at the tap. *Water Res.* 2012, 46, 5205–5216. [CrossRef] [PubMed]

272. DeSantis, M.K.; Triantafyllidou, S.; Schock, M.R.; Lytle, D.A. Mineralogical Evidence of Galvanic Corrosion in Drinking Water Lead Pipe Joints. *Environ. Sci. Technol.* 2018, 52, 3365–3374. [CrossRef] [PubMed]

273. Zhang, X.G. Chapter 10: Galvanic Corrosion. In *Uhlig’s Corrosion Handbook: Third Edition*; Revie, R.W., Ed.; The Electrochemical Society: Pennington, NJ, USA, 2011; pp. 123–143. ISBN 978-0-8031-0981-0.

274. Hack, H.P. Galvanic Corrosion—Google Books; American Society for Testing and Materials: Philadelphia, PA, USA, 1988.

275. Souto, R.M.; González-García, Y.; Bastos, A.C.; Simões, A.M. Investigating corrosion processes in the micrometric range: A SVET study of the galvanic corrosion of zinc coupled with iron. *Corros. Sci.* 2007, 49, 4568–4580. [CrossRef]

276. Marques, A.G.; Taryba, M.G.; Panão, A.S.; Lamaka, S.V.; Simões, A.M. Application of scanning electrode techniques for the evaluation of iron-zinc corrosion in nearly neutral chloride solutions. *Corros. Sci.* 2016, 104, 123–131. [CrossRef]

277. Kuntyi, O.I.; Zozulya, H.I.; Dobrovets’ka, O.Y.; Kornii, S.A.; Reshetnyak, O.V. Deposition of Copper, Silver, and Nickel on Aluminum by Galvanic Replacement. *Mater. Sci.* 2018, 53, 488–494. [CrossRef]

278. Jorcin, J.-B.; Blanc, C.; Pêbère, N.; Tribollet, B.; Vivier, V. Galvanic Coupling Between Pure Copper and Pure Aluminum. *J. Electrochem. Soc.* 2008, 155, C46. [CrossRef]

279. Clark, B.N.; Masters, S.V.; Edwards, M.A. Lead Release to Drinking Water from Galvanized Steel Pipe Coatings. *Environ. Eng. Sci.* 2015, 32, 713–721. [CrossRef]

280. Zhou, M.; Liu, C.; Xu, S.; Gao, Y.; Jiang, S. Accelerated degradation rate of AZ31 magnesium alloy by copper additions. *Mater. Corros.* 2018, 69, 760–769. [CrossRef]

281. Tada, E.; Sugawara, K.; Kaneko, H. Distribution of pH during galvanic corrosion of a Zn/steel couple. *Electrochim. Acta* 2004, 49, 1019–1026. [CrossRef]

282. United States Bureau of the Census. American Housing Survey, 2017 National Data. Available online: https://www.census.gov/programs-surveys/ahs/data/2017/ahs-2017-public-use-file--pubf-ahs-2017-national-public-use-file--pubf.html (accessed on 10 October 2020).

283. Ewann, F.; Hoffman, P.S. Cysteine Metabolism in Legionella pneumophila: Characterization of an L-Cystine-Utilizing Mutant. *Appl. Environ. Microbiol.* 2006, 72, 3993–4000. [CrossRef]

284. Shaheen, M.; Scott, C.; Ashbolt, N.J. Long-term persistence of infectious Legionella with free-living amoebae in drinking water biofilms. *Int. J. Hyg. Environ. Health* 2019, 222, 678–686. [CrossRef] [PubMed]

285. Richards, A.M.; Von Dwingelo, J.E.; Price, C.T.; Abu Kwaik, Y. Cellular microbiology and molecular ecology of Legionella–amoeba interaction. *Virulence* 2013, 4, 307–314. [CrossRef] [PubMed]
286. Michel, R.; Burghardt, H.; Bergmann, H. Acanthamoeba, naturally intracellularly infected with Pseudomonas aeruginosa, after their isolation from a microbiologically contaminated drinking water system in a hospital. *Zent. Hgy. Umweltmed.* 1995, 196, 532–544.

287. Ji, P.; Rhoads, W.J.; Edwards, M.A.; Pruden, A. Impact of water heater temperature setting and water use frequency on the building plumbing microbiome. *Isme J.* 2017, 11, 1318–1330. [CrossRef] [PubMed]

288. Buse, H.Y.; Lu, J.; Lu, X.; Mou, X.; Ashbolt, N.J. Microbial diversities (16S and 18S rRNA gene pyrosequencing) and environmental pathogens within drinking water biofilms grown on the common premise plumbing materials unplasticized polyvinylchloride and copper. *FEMS Microbiol. Ecol.* 2014, 88, 280–295. [CrossRef] [PubMed]

289. Falkinham, J.O.; Hilborn, E.D.; Arduino, M.J.; Pruden, A.; Edwards, M.A. Epidemiology and Ecology of Opportunistic Premise Plumbing Pathogens: Legionella pneumophila, Mycobacterium avium, and Pseudomonas aeruginosa. *Environ. Health Perspect.* 2015, 123, 749–758. [CrossRef] [PubMed]

290. Ma, X.; Pierre, D.; Bibby, K.; Stout, J.E. Bacterial community structure correlates with Legionella pneumophila colonization of New York City high rise building premises plumbing system. *Environ. Sci. Water Res. Technol.* 2020, 6, 1324–1335. [CrossRef]

291. Paranjape, K.; Bédard, É.; Whyte, L.G.; Ronholm, J.; Prévost, M.; Faucher, S.P. Presence of Legionella spp. in cooling towers: The role of microbial diversity, Pseudomonas, and continuous chlorine application. *Water Res.* 2020, 169, 115252. [CrossRef]

292. Garner, E.; McLain, J.; Bowers, J.; Engelthaler, D.M.; Edwards, M.A.; Pruden, A. Microbial Ecology and Water Chemistry Impact Regrowth of Opportunistic Pathogens in Full-Scale Reclaimed Water Distribution Systems. *Environ. Sci. Technol.* 2018, 52, 9056–9068. [CrossRef]

293. Ji, P.; Parks, J.; Edwards, M.A.; Pruden, A. Impact of Water Chemistry, Pipe Material and Stagnation on the Building Plumbing Microbiome. *PloS ONE* 2015, 10, e0141087. [CrossRef]

294. Lin, W.; Yu, Z.; Chen, X.; Liu, R.; Zhang, H. Molecular characterization of natural biofilms from household taps with different materials: PVC, stainless steel, and cast iron in drinking water distribution system. *Appl. Microbiol. Biotechnol.* 2012, 97. [CrossRef] [PubMed]

295. Proctor, C.R.; Gächter, M.; Kötzsch, S.; Rölli, F.; Sigrist, R.; Walser, J.-C.; Hammes, F. Biofilms in shower hoses–choice of pipe material influences bacterial growth and communities. *Environ. Sci. Water Res. Technol.* 2016, 2, 670–682. [CrossRef]

296. Corre, M.-H.; Delafont, V.; Legrand, A.; Berjeaud, J.-M.; Verdon, J. Exploiting the Richness of Environmental Waterborne Bacterial Species to Find Natural Legionella pneumophila Competitors. *Front. Microbiol.* 2019, 9, 3360. [CrossRef] [PubMed]

297. Kimura, S.; Tateda, K.; Ishii, Y.; Horikawa, M.; Miyairi, S.; Gotoh, N.; Ishiguro, M.; Yamaguchi, K. Pseudomonas aeruginosa Las quorum sensing autoinducer suppresses growth and biofilm production in Legionella species. *Microbiology* 2009, 155, 1934–1939. [CrossRef] [PubMed]

298. Declerck, P.; Behets, J.; Delaetd, Y.; Margineanu, A.; Lammertyn, E.; Ollevier, F. Impact of Non-Legionella Bacteria on the Uptake and Intracellular Replication of Legionella pneumophila in Acanthamoeba castellanii and Naegleria lovaniensis. *Microb. Ecol.* 2005, 50, 536–549. [CrossRef]

299. Berry, D.; Horn, M.; Xi, C.; Raskin, L. Mycobacterium avium Infections of Acanthamoeba Strains: Host Strain Variability, Grazing-Induced Infections, and Altered Dynamics of Inactivation with Monochloramine. *Appl. Environ. Microbiol.* 2010, 76, 6685–6688. [CrossRef]

300. Weitere, M.; Bergfeld, T.; Rice, S.A.; Matz, C.; Kjelleberg, S. Grazing resistance of Pseudomonas aeruginosa biofilms depends on type of protective mechanism, developmental stage and protozoan feeding mode. *Environ. Microbiol.* 2005, 7, 1593–1601. [CrossRef]

301. Matz, C.; Moreno, A.M.; Alhede, M.; Manefield, M.; Hauser, A.R.; Givskov, M.; Kjelleberg, S. Pseudomonas aeruginosa uses type III secretion system to kill biofilm-associated amoebae. *ISME J.* 2008, 2, 843–852. [CrossRef]

302. Steinberg, C.; Laurent, J.; Edel-Hermann, V.; Barbezant, M.; Sixt, N.; Dalle, F.; Aho, S.; Bonnin, A.; Hartemann, P.; Sautour, M. Adaptation of Fusarium oxysporum and Fusarium dimerum to the specific aquatic environment provided by the water systems of hospitals. *Water Res.* 2015, 76, 53–65. [CrossRef]

303. Petitjean, M.; Martak, D.; Silvant, A.; Bertrand, X.; Valot, B.; Hoquet, D. Genomic characterization of a local epidemic Pseudomonas aeruginosa reveals specific features of the widespread clone ST395. *Microb. Genom.* 2017, 3, e000129. [CrossRef]
304. Oliver, J.D. Recent findings on the viable but nonculturable state in pathogenic bacteria. *FEMS Microbiol. Rev.* 2010, 34, 415–425. [CrossRef] [PubMed]

305. Williams, K.; Pruden, A.; Falkingham, J.O.; Edwards, M. Relationship between Organic Carbon and Opportunistic Pathogens in Simulated Glass Water Heaters. *Pathogens* 2015, 4, 355–372. [CrossRef] [PubMed]

306. Dusserre, E.; Ginevra, C.; Hallier-Soulier, S.; Vandenesch, F.; Festoc, G.; Etienne, J.; Jarraud, S.; Molmeret, M. A PCR-based method for monitoring Legionella pneumophila in water samples detects viable but noncultivable legionellae that can recover their cultivability. *Appl. Environ. Microbiol.* 2008, 74, 4817–4824. [CrossRef] [PubMed]

307. Wang, H.; Bédard, E.; Prévost, M.; Camper, A.K.; Hill, V.R.; Pruden, A. Methodological approaches for monitoring opportunistic pathogens in premise plumbing: A review. *Water Res.* 2017, 117, 68–86. [CrossRef] [PubMed]

308. Rowland, J.L.; Niederweis, M. Resistance mechanisms of Mycobacterium tuberculosis against phagosomal copper overload. *Tuberculosis* 2012, 92, 202–210. [CrossRef]

309. Lehtola, M.J.; Miettinen, I.T.; Keinänen, M.M.; Kekki, T.K.; Laine, O.; Hirvonen, A.; Vartiainen, T.; Martikainen, P. Microbiology, chemistry and biofilm development in a pilot drinking water distribution system with copper and plastic pipes. *Water Res.* 2004, 38, 3769–3779. [CrossRef]

310. Schwan, W.R.; Warrenre, P.; Keunz, E.; Kendall Stover, C.; Folger, K.R. Mutations in the cueA gene encoding a copper homeostasis P-type ATPase reduce the pathogenicity of Pseudomonas aeruginosa in mice. *Int. J. Med. Microbiol.* 2005, 295, 237–242. [CrossRef]

311. Abdollahi, S.; Rasooli, I.; Mousavi Gargari, S.L. The role of TonB-dependent copper receptor in virulence of Acinetobacter baumannii. *PLoS ONE* 2013, 8, e575. [CrossRef] [PubMed]

312. Alquethamy, S.; Khorvash, M.; Pederick, V.; Whittall, J.; Paton, J.; Paulsen, I.; Hassan, K.; McDevitt, C.; Sado, M.J.; Iglewski, B.H.; Ives, S.K.; Sokol, P.A.; Woods, D.E. Relationship of iron and extracellular virulence factors to Pseudomonas aeruginosa lung infections. *J. Med. Microbiol.* 1999, 34, 399–413. [CrossRef] [PubMed]

313. Huston, W.M.; Jennings, M.P.; McEwan, A.G. The multicopper oxidase of Pseudomonas aeruginosa is a ferroxidase with a central role in iron acquisition. *Mol. Microbiol.* 2002, 45, 1741–1750. [CrossRef] [PubMed]

314. James, B.W.; Mauchline, W.S.; Fitzgeorge, R.B.; Dennis, P.J.; Keevil, C.W. Influence of iron-limited continuous culture on physiology and virulence of Legionella pneumophila. *Infect. Immun.* 1995, 63, 4224–4230. [CrossRef] [PubMed]

315. Pruden, A. Balancing Water Sustainability and Public Health Goals in the Face of Growing Concerns about Antibiotic Resistance. *Environ. Sci. Technol.* 2014, 48, 5–14. [CrossRef]

316. Vasil, M.L.; Ochsner, U.A. The response of Pseudomonas aeruginosa to iron: Genetics, biochemistry and virulence. *Microbiology* 1999, 145, 210–220. [CrossRef] [PubMed]

317. Metz, M.L.; Sellin Je, P.; Eijkelkamp, B. The Role of the CopA Copper Eflux System in Acinetobacter baumannii Virulence. *Mol. Sci.* 2019, 6, 295. [CrossRef]

318. Bjorn, M.J.; Iglewski, B.H.; Ives, S.K.; Sado, M.J.; Ives, S.K.; Sokol, P.A.; Woods, D.E. Relationship of iron and extracellular virulence factors to Pseudomonas aeruginosa lung infections. *J. Med. Microbiol.* 2012, 60, 399–413. [CrossRef] [PubMed]

319. Fiester, S.E.; Arivett, B.A.; Schmidt, R.E.; Beckett, A.C.; Ticak, T.; Carrier, M.V.; Ghosh, R.; Ohneck, E.J.; Metz, M.L.; Sellin Je, P.; Eijkelkamp, B. The Role of the CopA Copper Eflux System in Acinetobacter baumannii Virulence. *Mol. Sci.* 2019, 6, 295. [CrossRef] [PubMed]

320. Berg, J.; Tom-Petersen, A.; Nybroe, O. Copper amendment of agricultural soil selects for bacterial antibiotic resistance in the field. *Lett. Appl. Microbiol.* 2005, 40, 146–151. [CrossRef]

321. Hu, H.; Wang, J.-T.; Li, J.; Li, J.-J.; Ma, Y.-B.; Chen, D.; He, J.-Z. Field-based evidence for copper contamination induced changes of antibiotic resistance in agricultural soils. *Environ. Microbiol.* 2016, 18, 3896–3909. [CrossRef]

322. Pruden, A. Balancing Water Sustainability and Public Health Goals in the Face of Growing Concerns about Antibiotic Resistance. *Environ. Sci. Technol.* 2014, 48, 5–14. [CrossRef]

323. Hu, H.; Wang, J.-T.; Li, J.; Li, J.; Ma, Y.-B.; Chen, D.; He, J.-Z. Field-based evidence for copper contamination induced changes of antibiotic resistance in agricultural soils. *Environ. Microbiol.* 2016, 18, 3896–3909. [CrossRef]

324. Knapp, C.W.; McCluskey, S.M.; Singh, B.K.; Campbell, C.D.; Hudson, G.; Graham, D.W. Antibiotic Resistance Gene Abundances Correlate with Metal and Geochemical Conditions in Archived Scottish Soils. *PLoS ONE* 2011, 6, e27300. [CrossRef] [PubMed]
325. Berg, J.; Thorsen, M.K.; Holm, P.E.; Jensen, J.; Nybroe, O.; Brandt, K.K. Cu Exposure under Field Conditions Coselects for Antibiotic Resistance as Determined by a Novel Cultivation-Independent Bacterial Community Tolerance Assay. *Environ. Sci. Technol.* 2010, 44, 8724–8728. [CrossRef] [PubMed]

326. Graham, D.W.; Olivares-Rieumont, S.; Knapp, C.W.; Lima, L.; Werner, D.; Bowen, E. Antibiotic Resistance Gene Abundances Associated with Waste Discharges to the Almendares River near Havana, Cuba. *Environ. Sci. Technol.* 2011, 45, 418–424. [CrossRef] [PubMed]

327. Zhang, M.; Chen, L.; Ye, C.; Yu, X. Co-selection of antibiotic resistance via copper shock loading on bacteria from a drinking water bio-filter. *Environ. Pollut.* 2018, 233, 132–141. [CrossRef]

328. Becerra-Castro, C.; Machado, R.; Vaz-Moreira, I.; Manaia, C. Assessment of copper and zinc salts as selectors of antibiotic resistance in Gram-negative bacteria. *Sci. Total Environ.* 2015, 530. [CrossRef]

329. Barker, J.; Scaife, H.; Brown, M.R. Intraphagocytic growth induces an antibiotic-resistant phenotype of Legionella pneumophila. *Antimicrob. Agents Chemother.* 1995, 39, 2684. [CrossRef]

330. Oglesby-Sherrouse, A.G.; Djapgne, L.; Nguyen, A.T.; Vasil, A.I.; Vasil, M.L. The complex interplay of iron, biofilm formation, and mucoidy affecting antimicrobial resistance of *Pseudomonas aeruginosa*. *Pathog. Dis.* 2014, 70, 307–320. [CrossRef]

331. Guo, X.; Liu, S.; Wang, Z.; Zhang, X.; Li, M.; Wu, B. Metagenomic profiles and antibiotic resistance genes in gut microbiota of mice exposed to arsenic and iron. *Chemosphere* 2014, 112, 1–8. [CrossRef]

332. European Commission. Council Directive 98/83/EC of 3 November 1998 on the quality of water intended for human consumption. Available online: https://eur-lex.europa.eu/legal-content/EN/TXT/?uri=CELEX%3A31998L0083 (accessed on 10 November 2020).

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