Uncertainty of runoff sensitivity to climate change in the Amazon River basin

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We employ the approach of Roderick and Farquhar (2011) to assess the sensitivity of runoff ($R$) given changes in precipitation ($P$), potential evapotranspiration ($E_p$), and other properties that change the partitioning of $P$ (n) by estimating coefficients that predict the weight of each variable in the relative change of $R$. We use this framework using different data sources and products for $P$, actual evapotranspiration ($E$), and $E_p$ within the Amazon River basin to quantify the uncertainty of the hydrologic response at the subcatchment scale. We show that when estimating results from the different combinations of datasets for the entire river basin (at Óbidos), a 10% increase in $P$ would increase $R$ on average 16%, while a 10% increase in $E_p$ would decrease $R$ about 6%. In addition, a 10% change in the parameter $n$ would affect the hydrological response of the entire basin around 5%. However, results change from catchment to catchment and are dependent on the combination of datasets. Finally, results suggest that enhanced estimates of $E$ and $E_p$ are needed to improve our understanding of the future scenarios of hydrological sensitivity with implications for the quantification of climate change impacts at the regional (subcatchment and subbasin) scale in Amazonia.

Keywords: Budyko framework; runoff sensitivity; evapotranspiration; catchment hydrology; Amazon River basin

Introduction

The Amazon River basin (ARB) is one of the few environmental tipping points of the Earth’s system. In this context, the term tipping point is employed to define a critical threshold at which a small perturbation can qualitatively alter the state or development of a system.1–3 Indeed, there is a strong body of the literature showing the likely die-back of the Amazon rainforest owing to the positive feedbacks between reduction in rainfall, temperature increase, and deforestation.4,5 Such transition would not be gradual but sudden when certain critical thresholds are surpassed, such as a 4 °C increase in temperature or a 40% increase in deforestation. At those critical points, the forest ceases to behave as a carbon sink and becomes a source, such that the forest collapses and is replaced by a savanna-type vegetation.1–3,6–14 On the other hand, in terms of hydrological regulation,15 it shows that catchments in the ARB can shift from a regulated to a nonregulated state when deforestation surpasses 30–40% of the forest cover. Then, floods and droughts become exacerbated, therefore, amplifying the variance of hydrological extremes, owing to the loss of regulation capacity in the “forest reservoir.” Such a shift in the hydroclimatic and ecological functioning of the whole ARB could have long-term consequences for the climatic system worldwide since the hydrological, climatological, and biogeochemical dynamics.

doi: 10.1111/nyas.14515
Table 1. Summary of possible dataset combinations

| Dataset combination | Precipitation $P$ (mm yr$^{-1}$) | Actual evapotranspiration $E$ (mm yr$^{-1}$) | Potential evapotranspiration $E_p$ (mm yr$^{-1}$) | Maximum number of subcatchments |
|---------------------|----------------------------------|---------------------------------------------|-----------------------------------------------|--------------------------------|
| 1                   | HYBAM                            | MPI                                         | Hargreaves                                    | 146                            |
| 2                   | HYBAM                            | MPI                                         | Priestley and Taylor                          | 146                            |
| 3                   | HYBAM                            | GLEAM                                       | Hargreaves                                    | 146                            |
| 4                   | HYBAM                            | GLEAM                                       | Priestley and Taylor                          | 146                            |
| 5                   | GPCC                             | MPI                                         | Hargreaves                                    | 146                            |
| 6                   | GPCC                             | MPI                                         | Priestley and Taylor                          | 146                            |
| 7                   | GPCC                             | GLEAM                                       | Hargreaves                                    | 146                            |
| 8                   | GPCC                             | GLEAM                                       | Priestley and Taylor                          | 146                            |
| 9                   | HYBAM                            | P–R                                         | Hargreaves                                    | 100                            |
| 10                  | HYBAM                            | P–R                                         | Priestley and Taylor                          | 100                            |
| 11                  | GPCC                             | P–R                                         | Hargreaves                                    | 100                            |
| 12                  | GPCC                             | P–R                                         | Priestley and Taylor                          | 100                            |

play an important role in regulating water, energy, and carbon budgets not only in South America but also are key hydroclimatic drivers at continental and global scales.

Multiple studies have dealt with the estimation of the long-term mean surface water balance of the ARB since it provides a coherent framework to understand the spatiotemporal dynamics of hydrologic fluxes across river basins (see Table 1 of Ref. 16). In general, these studies have addressed the water balance for the entire river basin (at Óbidos gauging station) and few of them report near steady-state conditions or almost exact water balance closure. However, as pointed out by Builes-Jaramillo and Poveda,16 such results of perfect closure are mostly influenced by the dependence of datasets in estimations of evapotranspiration ($E$), that is, $E$ derived from precipitation ($P$) products. Another source for the lack of a perfect closure of the long-term water balance is the possible existence of trends in a series of rainfall and river flows that violate the mean value theorem, and thus $R \neq P - E$. Studies by Greve et al.17 and Moussa et al.18 have provided insights to deal with this problem. At smaller spatial scales in the ARB (subcatchment scale), calculations of the closure of the water balance present major discrepancies, especially between observed data and reanalysis datasets.16

Thus, estimation of hydrologic fluxes is dependent on the chosen datasets in terms of their quality, accuracy, record lengths, and the presence of trends and changes. These inconsistencies in water balance estimations affect runoff calculations and the quantification of the combined impacts from climate variability, anthropogenic climate change, and deforestation and land use changes, given the complex nature of hydroclimatological processes. Thus, a deeper understanding of the spatiotemporal dynamics of river basins is required. This analysis could include the coupling between water and energy budgets, considering that water cycle in Amazonia is sensitive to perturbations caused by human-related activities19 and it is also driven by diverse natural climate variability phenomena, including the El Niño–Southern Oscillation (ENSO)20–26 and changes in sea surface temperatures in the Tropical North Atlantic, among others.27–29

The theoretical framework introduced by Budyko provides a functional relationship to understand the coupling between water and energy budgets in river basins. It has been used to address fundamental questions in hydrology, such as water availability and management, runoff prediction, and climate change studies.17,30–43 It has also been used at various spatial and temporal scales to perform diagnostic analyses of the long-term mean annual water balances in catchments and to study the interactions between hydroclimate, soil, vegetation, and topography.44–50

This study aims to: (1) estimate the closure of the surface water balance across subcatchments of
the ARB comparing different datasets of \( P, E, \) and \( E_p \); (2) employ Budyko’s framework using Roderick and Farquhar’s approach\(^{35} \) to assess the sensitivity of runoff (\( R \)) in the ARB, given changes in climatic variables (\( P \)) and potential evapotranspiration (\( E_p \)), as well as in other properties that affect the partitioning of \( P \), represented by the parameter \( n \), by estimating the sensitivity coefficients that analytically predict the weight of each variable (\( P, E_p, \) and \( n \)) in the relative change of \( R \) across the ARB and its six major subbasins. In particular, we aim to (3) investigate the spatial variability of three sensitivity coefficients across the major subcatchments of the ARB and to (4) quantify the uncertainty in the estimation of the sensitivity coefficients given the currently available datasets on the water budget components of the ARB catchments.

Materials and methods

Study area and datasets

The ARB is the largest river basin in the world and constitutes a singular biogeographic, physiographic, hydrological, and climatic entity rich in biodiversity. It is located in tropical South America, between 5°N and 20°S, and 80°W and 50°W, approximately. It drains an area of about 6,700,000 km\(^2\) across important portions of Brazil, Bolivia, Colombia, Ecuador, Peru, Venezuela, Suriname, and French Guyana. Its main channel has a length of 6280 km with a mean annual flow of 209,000 m\(^3\) s\(^{-1}\) at its farthest downstream gauging station, Óbidos.\(^{31} \) This river basin can be divided into 146 subcatchments, which, in turn, can be grouped into six major river subbasins (Fig. 1): Solimões, Purus, Negro, Madeira,
There are approximately 15,000 tributaries in the ARB, including the Madeira River, which is considered the longest one (3380 km), and the Negro River with the largest discharge of up to 100,000 m$^3$ s$^{-1}$ (Ref. 52).

Since one of our goals is to estimate the closure of the surface water balance across the ARB and to quantify the uncertainty in the estimation of the sensitivity coefficients comparing different datasets, we used at least two sources of data for each hydrological variable, except for runoff. $P$ [LT$^{-1}$] and river discharges, $Q$ [LT$^{-3}$T$^{-1}$], were obtained from the Observation Service SO-HYBAM (formerly the Environmental Research Observatory ORE-HYBAM), available at http://www.ore-hybam.org/. Precipitation data were also used from the Global Precipitation Climatology Centre (GPCC), calculated from global station data. River discharge records, $Q$, were transformed into runoff, $R$ [LT$^{-1}$], considering that $R = Q/A$, with $A$ [L$^2$] being the drainage area of each subcatchment. Data regarding actual $E$ [LT$^{-1}$] were obtained from a global monitoring network combined with meteorological observations and remote sensing data, compiled by the Max Planck Institute for Biogeochemistry (MPI) as well as from the Global Land Evaporation Amsterdam Model (GLEAM), a set of algorithms that separately estimate the different components of land evapotranspiration. We also used actual $E$ estimated from the water balance equation ($E = P-R$) with data from precipitation (HYBAM and GPCC) and observed runoff. For $E_p$ [LT$^{-1}$], we also used two different methods. First, we used the often-applied Hargreaves equation:

$$E_p = 0.0023 \left( T_{\text{max}} - T_{\text{min}} \right)^{0.5} \left( T_{\text{mean}} + 17.8 \right) R_n, \quad (1)$$

where $T_{\text{max}}$, $T_{\text{min}}$, and $T_{\text{mean}}$ are the maximum, minimum, and mean air temperatures ($^\circ$C), respectively, and $R_n$ is the extraterrestrial radiation or solar radiation at the top of the atmosphere. We used this equation following Trabucco and Zomer who compared four different temperature-based methods for $E_p$ and concluded that for South America, this model is one of the most suitable ways to estimate $E_p$. The Hargreaves equation was estimated using station data of daily minimum and maximum temperatures from the Climatic Research Unit (CRU). Finally, $E_p$ was also estimated using the Priestley–Taylor equation:

$$E_p = \alpha \left( \frac{\Delta}{\Delta + \gamma} \right) (R_n - G), \quad (2)$$

with the Priestley–Taylor empirical parameter $\alpha = 1.26$, the slope of the saturation vapor pressure curve ($\Delta$) as a function of temperature, the psychrometric constant $\gamma$, net radiation $R_n$, and ground heat flux $G$. We used annual average data from the remote sensing product of the Surface Radiation Budget (SRB; https://eosweb.larc.nasa.gov/project/srb/srb_table), at 1-degree resolution for net radiation and surface temperature to estimate $\Delta$. The ground heat flux was neglected since we use annual mean data.

All variables ($P$, $R$, $E$, and $E_p$) were available at a monthly temporal resolution (mm month$^{-1}$) from January 1984 to December 2007, thus covering 24 years of data. Also, $P$ and $E$ were available for 146 subcatchments, while $Q$ was available only for 100 of them. For this reason, the analysis requiring calculations of runoff was only carried out for this subset of subcatchments. Table 1 summarizes the 12 possible combinations of datasets used in this study.

The water balance equation

In a now-classic paper, Dooge stated that the fundamental business of hydrology is to solve the water balance equation at a wide range of spatial and temporal scales. It is the differential equation representing the conservation of mass within a control volume defined by a river basin, and expressed as:

$$\frac{dS}{dt} = P(t) - R(t) - E(t), \quad (3)$$

where $S(t)$ represents water storage in soils and groundwater as a function of time, $P(t)$ and $E(t)$ denote, respectively, $P$ and actual $E$ in the river basin, and $R(t)$ denotes the surface runoff at the basin’s outlet. Integration of Eq. (3) over long time periods (15–20 years), the change in water storage with time (left-hand side) becomes negligible when compared with the other terms of the equation. In the long term, the right-hand side terms $P(t)$, $E(t)$, and $R(t)$ converge to their long-term mean values provided that the probability distribution function of the variables is stationary, according to the mean value theorem. Therefore, Eq. (3) becomes:

$$\hat{R} = \hat{P} - \hat{E}. \quad (4)$$
The terms in Eq. (4) have units of [LT$^{-1}$], and thus the river basin drainage area, $A$, becomes the scaling factor of runoff to obtain the long-term mean river streamflow, $Q$ [L$^3$T$^{-1}$] = $R$ [LT$^{-1}$] $A$ [L$^2$]. Hereinafter, we leave out the bars in Eq. (4) to denote the long-term annual mean values of the respective terms.

**The Budyko framework**

This theory relates actual $E$ with two variables representing the water and energy budgets, respectively: $P$ and $E_p$ for river basins over long timescales, and thus $E = f (P, E_p)$. Such a relationship is constrained by two physical limits: (1) $E \rightarrow E_p$ as $P \rightarrow \infty$ (energy-limited wet environments); and (2) $E \rightarrow P$ as $E_p \rightarrow \infty$ (water-limited dry environments). Budyko$^{62}$ went on to propose that $E/P = f(E_p/P)$, where the nondimensional variable $E_p/P$ is referred to as the aridity index ($\Phi$), and, therefore, the previous two limits become: (1) $E/P \rightarrow \Phi$ as $\Phi \rightarrow 0$ ($E$ is energy-limited); and (2) $E/P \rightarrow 1$ as $\Phi \rightarrow \infty$ ($E$ is water-limited). The ratio $E/P$ can be considered a measure of the long-term mean annual water balance in a river basin, since it is the fraction of the water falling as precipitation that is partitioned into evapotranspiration. On the other hand, $\Phi$ is a measure of the long-term mean climate-related energy availability to water availability. Small values of $\Phi$ ($\Phi < 1$) are associated with humid catchments where precipitation is significant, and the energy supply is the limiting factor for evapotranspiration. Conversely, large values of $\Phi$ ($\Phi > 1$) are found in arid regions where precipitation is low, and evapotranspiration is limited by water supply.

In this paper, we used the Budyko framework to study the relative change in runoff in a given catchment as a function of relative changes in two climatic variables ($P$ and $E_p$) and in other properties that affect the partitioning of $P$, represented by the parameter $n$. For this purpose, we employ the approach by Roderick and Farquhar$^{35}$ and the following equations (see Ref. 35 for their derivation):

$$\frac{dR}{R} = \left[ \frac{P}{R} \left(1 - \frac{\partial E}{\partial P}\right) \right] \frac{dP}{P} - \left[ \frac{E_p}{R} \frac{\partial E}{\partial E_p} \right] \frac{dE_p}{E_p}$$

$$- \left[ \frac{n \partial E}{R \partial n} \right] \frac{dn}{n}.$$  \hspace{1cm} (5)

The three terms in the brackets of Eq. (5) are referred to as runoff sensitivity coefficients, such that:

$$\varepsilon_p = \left[ \frac{P}{R} \left(1 - \frac{\partial E}{\partial P}\right) \right],$$

$$\varepsilon_{E_p} = \left[ \frac{E_p}{R} \frac{\partial E}{\partial E_p} \right],$$

$$\varepsilon_n = \left[ \frac{n \partial E}{R \partial n} \right].$$  \hspace{1cm} (6–8)

In these equations, the water balance constraint implies that the runoff sensitivity coefficients of $P$ and $E_p$ add up to one, as $\varepsilon_p + \varepsilon_{E_p} = 1$. Also, in Eqs. (5)–(8), $n$ represents other factors that change the partitioning of $P$ between $E$ and $R$. That is, $n$ captures all factors that are known to influence runoff generation, such as topography, soils, and vegetation and other aspects like changes in precipitation intensity or changes in the spatial distribution, and/or seasonal timing of $P$ and $E_p$. For example, Yang and Yang$^{66}$ suggested that $n$ increases in catchments with the amount of water available for $E$, thus, a large value of $n$ indicates that there is sufficient green water available for $E$ to achieve close to its maximum value under any climatic condition. For Eq. (8), $n$ was fitted for each subcatchment and each combination of datasets using interannual data and the Budyko-type equation by Yang $et al.$$^{67}$

The formulation by Roderick and Farquhar$^{35}$ assumes steady-state water balance conditions and, therefore, calculations require a timescale whereby changes in catchment storage are small relative to the magnitude of fluxes. For this reason, monthly values of $P$, $E$, $E_p$, and $R$ were converted to long-term means. Also, before any further analysis was carried out, we confirmed that the 12 possible combinations of datasets (mentioned above) follow the Budyko framework. This was done by plotting the location of each dataset in the Budyko space ($\Phi$ versus $E/P$). We found that not only this framework holds when using the proposed datasets but also most subcatchments are located in the humid space (where $\Phi < 1$), as it was thoroughly shown for the catchments of the ARB by Carmona $et al.$$^{39}$ These results are not shown here for the sake of brevity. We also calculated the Spearman’s rank correlation coefficient between $\Phi$ and $E/P$ for each dataset. This test
evaluates how well the relationship between two variables can be described using a monotonic function, with values from +1 (perfect positive association) to −1 (perfect negative association). The closer \( \rho \) is to zero, the weaker the association between the ranks. Results show moderate (+0.5) to strong (+0.7) correlations between our variables (\( E_p/P \) and \( E/P \)). This suggests that all combinations of datasets are useful for the analysis presented in this manuscript. However, according to these results, the best combinations are #1 (\( P_{HYBAM} - E_{MPI} - E_p \) Hargreaves, \( \rho = 0.95 \)) and #5 (\( P_{GPCC} - E_{MPI} - E_p \) Hargreaves, \( \rho = 0.96 \)), and the worst is #12 (\( P_{GPCC} - E = P - R - E_p \) Priestley–Taylor, \( \rho = 0.50 \)). We also found that among all, the weakest correlations are obtained when using \( E \) estimated from the water balance equation \( (P-R) \) as in combinations 9–12 in Table 1.

Finally, we follow the approach by Jung et al.\(^8\) to quantify the relative contribution of uncertainty. Therefore, we carried out a multilinear regression linking the differences between the ensembles of datasets for each subcatchment using the following equations:

\[
\Delta \epsilon_p = a^* \Delta P + b^* \Delta E + c^* \Delta E_p, \tag{9}
\]

\[
\Delta \epsilon_{E_p} = a^* \Delta P + b^* \Delta E + c^* \Delta E_p, \tag{10}
\]

\[
\Delta \epsilon_n = a^* \Delta P + b^* \Delta E + c^* \Delta E_p. \tag{11}
\]

In these equations, \( \Delta P, \Delta E_p, \) and \( \Delta E \) are the difference of each dataset of a variable \( (P, E, \) or \( E_p) \) and the mean value of all combinations of datasets for that variable. For example, \( \Delta P \) is the difference of each precipitation product \( (HYBAM \) or \( GPCC) \) and the mean value of \( P \) using all precipitation products.
(i.e., $\Delta P = P_{\text{HYBAM}} - P_{\text{mean}}$). Also, $\Delta \varepsilon_p$, $\Delta \varepsilon_{Ep}$, and $\Delta \varepsilon_n$ are the difference of each sensitivity coefficient (in Eqs. 6–8) of each dataset combination and the overall mean of the ensemble of 12 dataset combinations in Table 1. Equations (9)–(11) are used to estimate the relative contribution of each variable to each sensitivity coefficient in each subcatchment. Also, $a$, $b$, and $c$ are the coefficients of the multilinear regression.

**Results**

Recalling the objectives of the present study, we discuss below the spatial distribution of $P$, $E_p$, and $E_d$ and their variability across the ARB. The following subsection is focused on the estimation of the water balance closure using different measurements and estimations of $P$ and $E$ and observed $R$ to determine which combination of datasets yields the best runoff estimates. Finally, we estimate the sensitivity coefficients that analytically predict the weight of $P$, $E_p$, and $E_d$ in the relative change of $R$ across the ARB. We also explore their spatial variability across the major subcatchments and quantify the uncertainty in their calculation.

**Spatial variability of $P$**

Long-term mean annual rainfall fields estimated with HYBAM and GPCC data are shown in Figure 2 (panels A and B, respectively). For spatial comparison purposes, data from the Tropical Rainfall Measuring Mission (TRMM) for the period spanning from 2001 to 2009 were also plotted (Fig. 2C). The three maps show similar regional precipitation patterns, indicating that the long-term mean annual precipitation in the ARB ranges from 1100 mm yr$^{-1}$ to up to 3800 mm yr$^{-1}$, and that the highest annual precipitation intensities occur in the Colombian Amazon (northwestern region), while the lowest precipitation rates are witnessed in Peru (western region), Bolivia (southwestern region), and some parts of Brazil (southeastern region). To quantify the differences or similarities between the HYBAM and GPCC data, the relative difference between these datasets was estimated. These relative differences are of the order of 10% ($-10\%$ to $10\%$) in 75% of the 146 subcatchments, while in 25% subcatchments, this difference is higher than 10% (ranging from $-20\%$ to $66\%$).

**Spatial variability of $E_p$**

Figure 3 presents the long-term mean annual fields of $E_p$ estimated using the Hargreaves equation (panel A) with data from the CRU, and the Priestley–Taylor equation (panel B) using data from the SRB. Unlike for precipitation, this figure shows that both datasets produce very different spatial estimates, only coinciding in a few catchments located toward the center of the ARB. The Hargreaves method locates the highest $E_p$ values (up to 1885 mm yr$^{-1}$) over the eastern part of the ARB decreasing westward, while the Priestley–Taylor
estimates locate the highest $E_p$ values toward the northeast, decreasing systematically toward the southwest. It should be kept in mind that values in Figure 3A were calculated using temperature data and mean monthly extraterrestrial radiation, while values in Figure 3B were calculated using short- and longwave radiation. Estimated relative differences between datasets are about 10% (−10% to 10%) in 55% of the 146 subcatchments, while for 45% of them, the difference is higher than 10%, ranging from −20% to 26%.

**Spatial variability of actual $E$**

Figure 4 presents the long-term mean annual fields of actual $E$ obtained from MPI (Fig. 4A) and GLEAM (Fig. 4B). These two datasets produce quite different spatial patterns in the long-term mean annual actual $E$, although both coincide in that the region with the lowest values of $E$ located near Perú and Bolivia (Fig. 4A and B). For comparison purposes in those subcatchments where $Q$ data were available, $E$ was also estimated using the water balance equation as $E = P - R$ using precipitation data from HYBAM (Fig. 4C) and GPCC (Fig. 4D). Figure 4 shows that GLEAM estimates of $E$ exhibit a similar pattern to $E_p$ estimates using the Priestley–Taylor equation, showing the highest values toward the northeast. This is mostly because GLEAM also uses the Priestley–Taylor equation to calculate $E_p$ based on the observations of surface net radiation and near-surface air temperature. Also, Figure 4 shows that MPI estimates show less spatial variability compared with results using GLEAM or the water balance closure.

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Figure 4. Spatial distribution of long-term mean annual actual evapotranspiration ($E$) in the Amazon River basin with four different datasets: (A) MPI, (B) GLEAM, (C) difference between long-term mean $P$ from HYBAM and observed $R$, and (D) difference between long-term mean $P$ from GPCC and observed runoff.
Relative differences between datasets were estimated using $E$ from the water balance ($P$–$R$) as a reference, in order to compare results between MPI and GLEAM. These results show a possible disagreement between remote sensing actual $E$ and the water budget at the climatological mean. Results from the relative difference between $E$ from MPI and $E$ calculated with $P$ from HYBAM are among 10% in 55% of the subcatchments, while in 45% of catchments, these differences are higher than 10% (from $-36\%$ to $136\%$). For $E$ from GLEAM and $E$ with $P$ from HYBAM, in 48% of the subcatchments, differences are among 10%, while in 52% of subcatchments, differences are higher than 10% (from $-26\%$ to $154\%$). When using $E$ calculated with $P$ from GPCC, these percentages change as follows: (1) for $E$ from MPI, in 39% of the subcatchments, differences are among 10%, and in 61% of them, differences are higher than 10% (ranging from $-41\%$ to $238\%$), and (2) for $E$ from GLEAM, in 34% of the subcatchments, differences are among 10%, and in 66% of them, differences are higher than 10% (ranging from $-45\%$ to $210\%$). In general, differences seem to be smaller when using $E$ from MPI compared with $E$ from GLEAM. Also, relative differences increase when using $E$ calculated using $P$ data from GPCC.

**Closure of the water balance**

The water balance closure was examined for 100 subcatchments of the ARB in terms of the estimation of $R$, using four combinations of $P$ and $E$ datasets: (1) $P$ Hybam + $E$ MPI ($R_1$); (2) $P$ Hybam + $E$ Gleam ($R_2$); (3) $P$ GPCC + $E$ MPI ($R_3$); and (4) $P$ GPCC + $E$ Gleam ($R_4$). Then, the values of estimated $R = P – E$ were compared with the observed values of $R$. Figure 5 plots $R$ estimated using each combination of datasets with observed $R$. In this Figure, each dot represents one catchment. The coefficient of determination ($R^2$) and the root-mean-square error (RMSE) are used as a measure of how well observed outcomes are replicated by each model. As shown in Figure 5, results for the water balance closure of the 100 subcatchments indicate that the lowest RMSE values are obtained when calculating $R$ using $P$ from HYBAM with actual $E$ from MPI (RMSE = 220 mm yr$^{-1}$) and GLEAM (RMSE = 227 mm yr$^{-1}$). It is worth noticing in Figure 5 that high RMSE results are almost independent of the $E$ estimate but better with $P$ from HYBAM than from GPCC.

We also quantified deviations of $R$ catchment per catchment as the relative difference between estimated and observed runoff. Assuming we accept a 10% difference, for the entire ARB (at Óbidos), the
Table 2. Percentage of catchments with 10% or less than 10% difference between observed and estimated \( R \) using the four combinations of datasets

| Runoff models | RMSE (mm yr\(^{-1}\)) | Catchments with 10% (or less) difference between observed and estimated \( R \) |
|---------------|------------------------|---------------------------------------------------------------------|
| \( R_1 \) (HYB-MPI) | 220                    | 55%                                                                 |
| \( R_2 \) (HYB-GLEAM) | 227                    | 53%                                                                 |
| \( R_3 \) (GPCC-MPI) | 262                    | 44%                                                                 |
| \( R_4 \) (GPCC-GLEAM) | 275                    | 40%                                                                 |

Water balance closes better when using data from HYBAM and MPI (7%) or HYBAM and GLEAM (10%). For data using GPCC and MPI or GPCC and GLEAM, deviations are 11% and 14%, respectively. However, the closure of the long-term water budget is not entirely guaranteed with these combinations of datasets for all the subcatchments, since only for 40% (with \( R_4 \)) and up to 55% (with \( R_1 \)) of them we could say that the water balance does close or that the temporal changes in storage can be considered negligible. This is shown in Table 2 and Figure 6, which presents the frequency of the relative differences and their ranges. This figure shows that for most subcatchments, runoff is better estimated using \( P \) from HYBAM and actual \( E \) from MPI. Also, there are seven subcatchments with imbalances equal to 80% or up. These imbalances seem to be due to great differences between precipitation datasets. From these seven subcatchments, four belong to the subbasin Madeira (south of the ARB) and three are located northeast, close to the mouth of the main Amazon River. For example, in the subcatchment Serra Do Moa, \( P \) from HYBAM is 1440 mm yr\(^{-1}\), while \( P \) from GPCC is 2395 mm yr\(^{-1}\); and for the subbasin Madeira (at Fazenda Vista Alegre), \( P \) from HYBAM is 1862 mm yr\(^{-1}\), while \( P \) from GPCC is 2429 mm yr\(^{-1}\). Given the location of the subcatchments (south of the ARB) and their runoff data, it seems that the GPCC dataset is overestimating \( P \) since both have low observed \( R \).

Figure 7 shows the spatial distribution of runoff for each combination of datasets. Results show that regardless of the used model, all maps show a similar spatial pattern with the highest (lowest) runoff located in the northeastern (southern) region of the ARB. These results coincide with maps of precipitation (Fig. 2), indicating that regional runoff patterns in the ARB are driven mostly by precipitation distribution in the basin.

Results for the runoff sensitivity coefficients

Here, we perform an in-depth analysis of the runoff sensitivity coefficients in Eq. (5) and their constitutive terms (Eqs. 6–8) for each of the 100 subcatchments with runoff data, with the aim of quantifying relative changes in \( R \) as a function of relative changes in climate (\( P \) and \( E_p \)) and in the parameter \( n \), using 12 different combinations of \( P \), \( E \), and \( E_p \) datasets in Table 1.

Table 3 shows the minimum, maximum, and mean values of the sensitivity coefficients calculated using long-term data for the ARB at its furthermore gauging station, Óbidos, and for the main six subbasins: Solimões, Negro, Purus, Madeira, Xingú, and Tapajós (Fig. 1). It also shows the standard deviation (SD) as a measure of the amount of variation or dispersion between 12 different combinations of datasets.

In summary, our results suggest that depending on the dataset combination, at Óbidos, a 10% increase in \( P \) would increase \( R \) between 15% and 17%, while a 10% increase in \( E_p \) would decrease \( R \) between 5% and 7%. In addition, a 10% change in factors embedded in the parameter \( n \) would affect the entire basin between 5% and 6%. Note that results presented in Table 3 evidence the water balance constraint since \( \varepsilon_p \) is in all cases a unit higher than \( \varepsilon_{E_p} \) and thus, \( \text{SD}\varepsilon_p = \text{SD}\varepsilon_{E_p} \).

Comparison between the six major subbasins in Table 3 shows that Madeira exhibits the highest differences between datasets, evidenced in its SD. In this subbasin, a 10% increase in \( P \) would increase \( R \) between 15% and 33%, a 10% increase in \( E_p \) would decrease \( R \) between 5% and 23%, and 10% change in the parameter \( n \) would affect the entire basin...
between 2% and 7%. This subbasin was identified to be one with the highest imbalance in runoff estimation depending on the dataset.

Results at a smaller spatial scale (subcatchment scale) are presented in Figure 8. This figure shows the spatial variability of the mean value of the sensitivity coefficients (Fig. 8A and B) and their respective SD (Fig. 8C and D) for all the 100 subcatchments. Only maps for $\varepsilon_P$ and $\varepsilon_n$ are plotted given the water balance constraint between $\varepsilon_P$ and $\varepsilon_E$, evidenced in Table 3. In Figure 8A and B, darker colors represent regions with higher values of both $\varepsilon_P$ and $\varepsilon_n$. Thus, these regions are subcatchments in which changes in $P$ (Fig. 8A) or in other factors that affect the partitioning of $P$ (Fig. 8B) would affect runoff the most. Panels 8A and D show those regions in which there is a higher (lower) deviation from the results estimated using the different datasets, represented by their SD. Thus, those subcatchments have the highest uncertainty in the results. Regarding SD of $\varepsilon_P$, most of these subcatchments also belong to Madeira, located south of the ARB, coinciding with imbalances in water balance closure. As for uncertainty in $\varepsilon_n$, again higher values are evidenced toward the south (in Madeira) but also to the west near the Andes and to the northeast close to the mouth of the ARB.

**Input data uncertainty of streamflow sensitivity**

Next, we disentangle the different impacts of data uncertainty on the sensitivity coefficients $\varepsilon_P$ and $\varepsilon_n$ using Eqs. (9) and (11). We only show results for $\varepsilon_P$ and $\varepsilon_n$, since results for $\varepsilon_P$ and $\varepsilon_E$ are the same given the aforementioned water balance constraint evidenced for the Amazon at Óbidos and the six main river basins in Table 2. Table 4 shows results of the relative contributions of $P$, $E$, and $E_p$ to the sensitivity coefficients for the entire river basin (at Óbidos) and within the six main subbasins.

Table 4 shows that for the entire Amazon and its major subbasins, the relative contribution of actual $E$ to the sensitivity of $P$ is higher than the respective relative contribution of $P$ and $E_p$. Also, regarding results for the sensitivity of factors embedded in $n$, $\varepsilon_n$. Table 4 shows that for the entire Amazon at Óbidos, the relative contribution of $E_p$ to $\varepsilon_n$ is higher than the contribution of $P$ and $E$ to $\varepsilon_n$. However, results vary from catchment to catchment. Spatial variability of the relative contribution of each variable to $\varepsilon_P$ and $\varepsilon_n$ for 100 subcatchments in the ARB is shown in Figure 9. This figure shows that at the subcatchment scale, relative contributions of $E$ to $\varepsilon_P$ (Fig. 9B) dominate the picture in the ARB, followed by the relative contribution of $P$ to $\varepsilon_P$ (Fig. 9A). On
the other hand, in 62% of the subcatchments, the relative contribution of $E_p$ to $\varepsilon_n$ (Fig. 9F) is higher than the relative contribution of $P$ to $\varepsilon_n$ and $E$ to $\varepsilon_n$.

**Discussion**

In this study, we analyze the implications of using different datasets and products to assess the relative changes in runoff $R$ in the ARB, given relative changes in climatic variables ($P$ and $E_p$), as well as in other properties that affect the partitioning of $P$, represented by $n$.

First, we study the variability of different datasets and products of $P$, actual $E$, and $E_p$ and their regional distribution across the ARB. We show that all precipitation datasets exhibit similar regional patterns of $P$ with the highest values of annual precipitation occurring in the Colombian Amazon (northwestern region), while the lowest precipitation rates are depicted in Peru (western region), Bolivia (southwestern region), and some parts of Brazil (southeastern region). These observations coincide with macroclimatic factors, such as the migration of the Intertropical Convergence Zone (ITZC) and the South Atlantic Convergence Zone (SACZ), the activity of aerial rivers, and large-scale circulation patterns across South America, including the Eastern Andean jet, also known as the Orinoco jet (northernmost leg of the SALJET), and the two phases of the ENSO, land surface–atmosphere interactions, vegetation activity, and precipitation recycling. Regarding $E_p$, we show that both $E_p$ products produce quite different spatial results. The Hargreaves method locates the highest $E_p$ values (of up to 1885 mm yr$^{-1}$) over the eastern part of the ARB decreasing westward, while the Priestley–Taylor estimates

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**Figure 7.** Comparison of spatial patterns of estimated $R$ using four combinations of datasets of $P$, $E$, and observed $R$. Observed $R$ is the middle of the figure (A) allowing for comparison with $R1$ (B), $R2$ (C), $R3$ (D), and $R4$ (E).
Table 3. Results for the sensitivity coefficients in the Amazon River basin and its six major river basins

|              | Amazon (Óbidos) | Negro | Purus | Solimões | Madeira | Tapajós | Xingú |
|--------------|-----------------|-------|-------|----------|---------|---------|-------|
| ε<sub>P</sub>| Minimum         | 1.51  | 1.40  | 1.35     | 1.40    | 1.51    | 1.66  | 1.85  |
|              | Maximum         | 1.74  | 1.73  | 1.62     | 1.56    | 3.32    | 2.19  | 2.36  |
|              | Mean            | 1.64  | 1.56  | 1.50     | 1.49    | 2.06    | 1.83  | 2.04  |
|              | SD              | 0.08  | 0.10  | 0.10     | 0.06    | 0.62    | 0.17  | 0.19  |
| ε<sub>E</sub>| Minimum         | 0.51  | 0.40  | 0.35     | 0.40    | 0.51    | 0.66  | 0.85  |
|              | Maximum         | 0.74  | 0.73  | 0.62     | 0.56    | 2.32    | 1.19  | 1.36  |
|              | Mean            | 0.64  | 0.56  | 0.50     | 0.49    | 1.06    | 0.83  | 1.04  |
|              | SD              | 0.08  | 0.10  | 0.10     | 0.06    | 0.62    | 0.17  | 0.19  |
| ε<sub>n</sub>| Minimum         | 0.50  | 0.45  | 0.44     | 0.40    | 0.17    | 0.49  | 0.70  |
|              | Maximum         | 0.58  | 0.54  | 0.47     | 0.47    | 0.70    | 0.68  | 0.86  |
|              | Mean            | 0.53  | 0.49  | 0.46     | 0.44    | 0.52    | 0.58  | 0.74  |
|              | SD              | 0.02  | 0.03  | 0.01     | 0.03    | 0.20    | 0.07  | 0.05  |

locate the highest $E_p$ values toward the northeast, decreasing systematically toward the southwest. The key difference is that we use remote sensing–based observations of surface net radiation for the Priestley–Taylor based estimate, whereas the often-used Hargreaves equation is solely based on daily average and diurnal temperature range data and extraterrestrial radiation. Thereby, the Hargreaves equation contains empirical coefficients that have been established at local scales and these may not be constant throughout the region. As for the results for actual $E$, the two datasets used, namely, MPI and GLEAM, produce different spatial patterns in the long-term mean annual actual $E$, and although both coincide in that the region with the lowest values of $E$ appear to be near Perú and Bolivia. For comparison purposes in those subcatchments where $Q$ data were available, $E$ was also estimated using the water balance equation as $E = P - R$. Results show disagreement between remote sensing actual $E$ and the water budget at the climatological mean.

Above, we attempted to estimate the closure of the water balance. Our results allow us to conclude that the closure of the long-term water budget is not entirely guaranteed with these combinations of datasets for all the subcatchments. Assuming that we accept a 10% difference between estimated and observed values of $R$ for each subcatchment, for 40% and up to 55% of the subcatchments (depending on the combination of datasets), we could say that the water balance does close or that the changes in storage can be considered negligible. In the worst cases, these imbalances seemed to be driven by but not limited to differences in precipitation datasets. However, recalling the subsection “Spatial availability of actual $E$,” there is a large discrepancy in climatological water budget–derived $E$ and remote sensing–based estimates, since in about 45% of subcatchments, there are high relative differences between actual evapotranspiration datasets. Closing the water balance at a regional scale is most important to improve water resource assessment and its vulnerability to global change. However, it should be kept in mind that the water balance in a catchment may not close when all assumptions in Eq. (4) are not met. These assumptions include the requirement of long-term, high-quality hydrological measurements (over 20–30 years of data), negligible water storage changes from seasonal to interannual timescales, as well as contributions of groundwater net transport; and also, the stationary probability density functions of hydrological variables according to the mean value theorem. For example, Miguez-Macho and Fan, and Pokhrel et al. explore the role of groundwater in the Amazon water cycle, finding large contributions of lateral groundwater exchange, while Frappart et al. discuss large seasonal water storage, as derived from GRACE estimates in the northwestern region of the ARB. Also, historical trends in hydrological variables have been found in Amazonia. Data show increasing trends in temperature records, while precipitation shows mixed evidence of increasing and
Figure 8. Spatial distribution of the mean value and the standard deviation (SD) of the sensitivity coefficients: sensitivity to $P$ (A and C) and to $n$ (B and D), respectively.

decreasing trends throughout the ARB$^{89,90}$. Similarly, the mean and minimum river discharges exhibit decreasing trends in the Tapajós River, the upstream Madeira, and the Peruvian Amazon rivers. By contrast, increasing trends have been detected in mean and maximum river discharges in the northwestern Putumayo and Napo Rivers.$^{91}$ Thus, these possible signs of climate change in the ARB evidenced in trends in the long-term series might violate the assumption of stationarity.

Later on, following Roderick and Farquhar,$^{35}$ we quantified changes in $R$ as a function of changes in climate ($P$ and $E_p$) and in other factors represented by $n$, using the different combination of datasets of $E$ and $E_p$. It is important to recall that the parameter $n$ encompasses all factors that change the partitioning of $P$ between evaporation and stream-flow. Thus, it includes other climate characteristics, such as precipitation seasonality, timing, intensity, and form.$^{35,44,92–94}$ Specifically, as pointed out by Berghuijs and Woods,$^{94}$ disregarding the role of climate intraannual variability can bias the attribution of the parameter $n$ toward landscape properties, prevents landscape effects from being strictly separated from intraannual climate effects, overestimates the importance of landscape effects, and ignores the role of a part of the climate effects on water yield.

Our results show that at Óbidos, the furthest downstream gauging station along the Amazon River, theoretical results suggest that depending on the dataset combination, a 10% increase in $P$ would increase $R$ between 15% and 17%, while a 10% increase in $E_p$ would decrease $R$ between 5% and 7%. In addition, a 10% change in $n$ would affect
the entire basin between 5% and 6%. Also, between the six major subbasins, Madeira exhibits the highest uncertainty in results, which could be related to its water closure imbalance identified above. Other regions with high differences between datasets in the sensitivity in catchment properties are located west near the Andes and to the northeast close to the mouth of the ARB.

Figure 9. Relative contribution of $P$, $E$, and $E_p$ to the sensitivity coefficients $\varepsilon_p$ (maps on the left: A–C) and $\varepsilon_n$ (maps on the right: D–F).
Finally, when estimating the relative contribution of each input variable to $\epsilon_p$, and $\epsilon_n$, for 100 subcatchments in the ARB, results show that uncertainty of the sensitivity to climatic changes ($\epsilon_p$) is dominated by the uncertainty in the catchment scale $E$. By contrast, the uncertainty of sensitivity of streamflow to changes in catchment properties $\epsilon_n$ is dominated by atmospheric demand and supply.

Our results show that in most regions of the ARB, either $E$ or $E_p$, which showed most differences between datasets, dominates contributions in runoff sensitivity coefficients. On the other hand, there are also catchments in which uncertainty is due to differences in precipitation products, evidenced in the imbalance of the water balance closure. Thus, we show that using diverse combinations of datasets can lead to different results with implications for the quantification of climate change impacts at the sub-catchment scale in the ARB. Specifically, our results suggest that in order to improve our understanding of the possible future scenarios of hydrological sensitivity, enhanced estimates of $E$ and $E_p$ are needed since uncertainties in $E$ seem to dominate over the uncertainties in $P$ in the ARB.

**Acknowledgments**

We thank the Green Talents Programme, sponsored by the BMBF, which supported A.C.’s research stay at the Max-Planck Institute for Biogeochemistry where this work was done. A.K. and M.R. were funded by the Max-Planck Society. G.P. was supported by the DAAD and Universidad Nacional de Colombia at Medellín, Colombia.

Open access funding enabled and organized by Projekt DEAL.

**Competing interests**

The authors declare no competing interests.

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**Table 4. Relative contributions of $P$, $E$, and $E_p$ for the sensitivity coefficients of the six major river basins within the ARB**

| River basin       | $P$ to $\epsilon_p$ (%) | $E$ to $\epsilon_p$ (%) | $E_p$ to $\epsilon_p$ (%) | $P$ to $\epsilon_n$ (%) | $E$ to $\epsilon_n$ (%) | $E_p$ to $\epsilon_n$ (%) |
|-------------------|-------------------------|--------------------------|---------------------------|-------------------------|-------------------------|---------------------------|
| Amazon (Óbidos)   | 3.4                     | 91.26                    | 2.69                      | 3.16                    | 26.19                   | 58.49                     |
| Negro             | 10.41                   | 103.82                   | 1.86                      | 24.43                   | 12.31                   | 51.69                     |
| Purus             | 1.77                    | 100.84                   | 0.02                      | 69.16                   | 14.39                   | 0.00                      |
| Solimões          | 36.52                   | 138.67                   | 0.07                      | 85.10                   | 0.83                    | 0.06                      |
| Madeira           | 6.49                    | 112.89                   | 0.01                      | 43.60                   | 27.52                   | 0.00                      |
| Tapajós           | 15.05                   | 113.59                   | 3.01                      | 61.76                   | 1.47                    | 41.35                     |
| Xingu             | 4.89                    | 103.96                   | 0.18                      | 27.41                   | 31.60                   | 10.01                     |

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