Population Genetic Structure of Raccoons as a Consequence of Multiple Introductions and Range Expansion in the Boso Peninsula, Japan

Miki Hirose (miki.hirose.0317@gmail.com)  
Toho University

Kazuya Yoshida  
Toho University

Eiji Inoue  
Toho University

Masami Hasegawa  
Toho University

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Abstract

Raccoon (*Procyon lotor*) is a globally introduced invasive carnivore. Although controlling feral raccoon populations is important to reduce serious threats to local ecosystems, raccoons are not under rigid population control in Europe and Japan. We examined the D-loop and nuclear microsatellite regions to identify spatially explicit and feasible management units for effective population control and further range expansion retardation. Through the identification of five mitochondrial DNA haplotypes and three nuclear genetic groups, we identified at least three independent introductions, range expansion, and subsequent genetic admixture in the Boso Peninsula. Admitting that the currently recognizable two genetic clusters can be treated as different management units, these management units will soon fuse to a single but large population to which the effective population control will no longer be applicable due to the absence of a genetic barrier between southern and northern Chiba Prefecture.

Introduction

Controlling invasive species is important for reducing threats to biodiversity and natural resources on a global scale (Millennium Ecosystem Assessment 2005), because introduced species may disrupt local ecosystems and eliminate native species through predation and herbivory competition, pathogen transmission, or hybridization (Parker et al. 1999). Invasion may ultimately lead to serious human health (Crowl et al. 2008; Mazza et al. 2014) and economic (Pimentel et al. 2005; Vilà et al. 2010) problems.

Implementing biological control for invasive species often requires both systematic measures and adaptive management approaches (Lindenmayer and Likens 2009). For example, a systematic measure includes early detection and rapid response to eradicate newly established populations; monitoring population dynamics and range expansion are important components of adaptive management approaches. Not only distribution data but also molecular analyses are effective tools for understanding the process of range expansion by understanding the degree of genetic exchange under various environmental conditions (Lawson Handley et al. 2011). Moreover, an eradication program can be adopted by genetically identifying the marginal populations to limit range expansion and avoid genetic admixture (Fischer et al. 2017), whereas the core populations must be kept at low density to prevent negative impacts (Salgado 2018). Asada (2013) proposed "lag-phase management" as an effective management technique for mammals, considering the sexual difference in dispersal distance, as males disperse over significantly longer distances than females. When the target species have gender differences in dispersion distance, only males are distributed in the foremost part of the distribution expansion area, and a low-density region "lag-phase" is formed, where the Allee-effect (Gascoigne et al. 2009) retards population growth. Thereafter, the female spreads its distribution similar to "petals on a rose (Porter et al. 1991)," and an "increase-phase" is formed, where the number of animals increases explosively. Although it is difficult to achieve eradication, maintaining lag-phase by applying capture pressure is possible. Management in the invaded range, however, is not usually effective because of insufficient knowledge of the genetic structure associated with population ecology outside the native range (Salgado 2018).
Raccoon (*Procyon lotor*) is a carnivore originally native to North and Central America (Long 2003; Gehrt 2003). Due to its invasiveness in terms of wide dietary niche, habitat generalism, high density, and rapid population growth, feral raccoons are a serious threat to local ecosystems and eliminate native species through predation, competition, and pathogen transmission (Parker et al. 1999). Despite the urgency and necessity of raccoon population management, they are not under rigid population control in not only Europe (Salgado 2018), but also Japan (Ikeda 2004), because of increasing population trends, unlimited range expansion, and an inefficient management strategy (Ministry of the Environment Government of Japan. 2018; Salgado 2018).

To address the weakness of current raccoon population management strategy, we tried to understand the process of identifying spatially explicit and feasible management units for retarding population growth and further range expansion. We explored molecular genetic methods for understanding the process of range expansion and degree of genetic exchange in various landscapes (Lawson Handley et al. 2011) in a geographically limited peninsular region of Japan (Boso Peninsula of Chiba Prefecture). Since the relationship between raccoon expansion process and genetic exchange in Japan might be extremely complex (Okuyama et al. 2020), it is difficult to estimate the details of range expansion only from sighting records. However, if we focus on a geographically limited area, we might be able to examine the details of range expansion in association with landscape and genetic structures within a particular region.

In this study, we analyzed the mitochondrial D-loop and microsatellite regions of raccoons in the Chiba Prefecture to detect the area of introduced populations and expansion process, supplemented with information on past distribution (Ochiai et al. 2002, Asada 2014). Similar to those in Europe and Hokkaido (Okuyama et al. 2020; Salgado 2018), the feral raccoons in Chiba were estimated to be established from multiple founder events of pet escape or abandonment (Ochiai et al. 2002, Yoshida et al. 2020). We have discussed the recognition of management units for feral raccoons in a semi-confined landscape of the Boso Peninsula as a model system.

**Materials And Methods**

The area of Chiba Prefecture almost coincides with the Boso Peninsula (ca.2800 km²), protruding into the Pacific Ocean from the Kanto Plain (Fig. 1). The Kanto Plain, the largest plain surrounding the Tokyo metropolitan area, developed during the late Quaternary period through tectonic activities and glacio-eustatic sea-level changes (Sugai et al., 2013). The northern half of the Boso Peninsula is bordered by the Edogawa River along the northwestern margin and the Tonegawa River along the northwest to northeast margin, inclined northwestward from an altitude of 100 m in the southeast to 10 m in the northwest, because the central area of the Kanto Plain is still subsiding. In contrast, the southern part of the Boso Peninsula primarily consists of ancient but still uplifting terrain with a well-dissected steep valley, although the entire terrain is usually below 300 m altitude. These hilly terrains are intervened by the narrow coastal diluvial plain and are ultimately surrounded by the Tokyo Bay and the Pacific Ocean. The northwestern and coastal bay areas of the peninsula are located within the highly developed and densely
populated Tokyo metropolitan area, but the eastern and southern hilly regions are outskirts of the metropolitan area with sufficient cropland, grassland, and forest areas (Bagan and Yamagata 2012). The climate of the peninsula is warm–temperate, with 5.2–29.0 °C mean monthly temperature and 1193–2203 mm annual precipitation, as reported in 2020 (Japan Meteorological Agency).

Tissue samples were collected from 179 carcasses of feral racoons that were euthanized for pest control in Chiba Prefecture from November 2014 to August 2019. Mitochondrial DNA (mtDNA) haplotype and nuclear microsatellite loci were determined for the tail tissue preserved in 99.5% ethanol or buccal cells collected with a cotton swab, rinsed in phosphate-buffered saline (PBS) solution (137 mmol/L NaCl, 8.1 mmol/L Na₂HPO₄, 2.68 mmol/L KCl, and 1.47 mM KH₂PO₄), and preserved in medium with 70% concentration of ethanol until DNA extraction.

**DNA extraction and laboratory procedures**

DNA was extracted using DNeasy Blood & Tissue Kit (QIAGEN; Tokyo, Japan). The tissue samples were cut into small pieces and DNA was extracted according to the manufacturer's instructions. The buccal cell sample was rinsed three times in PBS solution and DNA was extracted. The DNA was eluted using 200 µL AE buffer.

A part of the mitochondrial D-loop region was amplified using PLO-L15997 (5′-CCATCAGCACCCAAAGCT-3′, Frantz et al. 2013) and PLO-CRL1 (5′-CGCTTAAACTTATGTCCTGTAACC-3′, Cullingham et al. 2008) primers. The PCR amplifications were performed in 15 µL total volume containing 1 µL template, 0.5 U Expand High Fidelity Enzyme Mix (Roche; Tokyo, Japan), 150 µM of each dNTP, and 300 nM of each primer. After an initial incubation at 95 °C for 5 min, PCR was performed for 35–40 cycles of denaturation at 94 °C for 30 s, annealing at 56 °C for 30 s, and extension at 72 °C for 30 s, followed by a final extension at 72 °C for 30 min. The PCR products were purified using EXoSAP-IT (Affymetrix; Cleveland, OH, USA) and sequenced using a Big Dye Terminator v3.1 Cycle Sequencing Kit (Thermo Fisher Scientific; Waltham, MA, USA) using the FASMAC sequencing service (FASMAC; Kanagawa, Japan). The sequence was aligned using MEGA 10 software (Kumar et al. 2018).

A total of 24 microsatellite loci (Cullingham et al. 2006, Fike et al. 2007, Siripunkaw et al. 2008) were analyzed via multiplex PCR (Table 2). To modify the amplicon length, new primers for three loci (PLO-M3, PLO-M17, and PLO-M15) were designed based on the sequences registered in GenBank using Primer 3 (Koressaar & Remm 2007, Untergasser et al. 2012; Table 1). We added a GTTCTT sequence to the 5′ end of the reverse primers of the loci that included dinucleotide repeats (Fike et al. 2007, Siripunkaw et al. 2008); this sequence promotes near complete adenylation of the 3′ end and decreases genotyping error risk (Browstein et al. 1996, Pompanon et al. 2005). We designed three multiplex PCR sets (Multiplex 1: 7 primers, Multiplex 2: 10 primers, Multiplex 3: 7 primers) using Multiplex Manager v1.2 (Holleley & Geerts 2009; Table 2). We added 1 µl mold DNA template and 2.5 µl Multiplex PCR Master Mix (QIAGEN) to a 5 µl mixture for performing the PCR. The concentration of each primer is listed in Table 2. The amplification conditions were as follows: 15 min at 95 °C, followed by 40 cycles of 94 °C for 30 s, 57 °C for 30 s, 72 °C
for 30 s, and a final extension at 60 °C for 30 min. We increased the number of cycles to 45 for the samples that did not yield adequate products after 40 PCR cycles. Fragment analysis of the PCR products was performed using the FASMAC fragment analysis service (Kanagawa, Japan) and the genotypes were determined using Peak Scanner Software v1.0 (Thermo Fisher Scientific K.K.; Tokyo, Japan).

Genotyping mtDNA

For the mtDNA control region, 406 bp sequences that overlapped with the same region used in a previous study (Yoshida et al. 2020) were determined, and a median-joining network, which included the published sequences in the native population (Cullingham et al. 2008) and 15 sequences from invasive raccoons in Japan retrieved from GenBank (NC_009126, AB297804, AB361247, AB462045-49, LC455747-53), was constructed using Network 5 (http://www.fluxus-engineering.com, Bandelt et al. 1999). We followed the categorization of lineages in the network described by Cullingham et al. (2008). Thereafter, the genetic diversity (He) of mtDNA was compared among the feral raccoons from Chiba, Hokkaido (Okuyama et al. 2020), those from Spain, Germany, Czech Republic, and Poland (Biedryzycka et al. 2014, Fischer et al. 2015, Alda et al. 2013), and the native raccoons from North America, including Ontario, Indiana, Columbia, Missouri, Illinois, and Missouri (Cullingham et al. 2006, Fike et al. 2007, Siripunkaw et al. 2008, Santonastaso et al. 2012, Alda et al. 2013) with a sigh test using R version 3.6.2. Thus, we downloaded the Excel file with geographical coordinates of each animal and the allelic fragment length at every gene locus from the Dryad database (http://datadryad.org/; doi:10.5061/dryad.412h3, Fischer et al. 2015). Information regarding the genetic diversity of raccoons of Germany (Fischer et al. 2015) was calculated using GenAlEx 6.502.

Genotyping and identification of genetic groups with microsatellite loci

Genetic diversity and differentiation among district groups of the raccoons were described initially by genotyping 179 raccoons sampled from the various localities of Chiba Prefecture, and by calculating the number of alleles (Na) in each locus, the effective number of alleles (Ne), observed heterozygosity (Ho), expected heterozygosity (He), private alleles (pA), and allelic richness using GenAlEx 6.502 (Peakall & Smouse 2006; Peakall & Smouse 2012) and FSTAT 2.9.3.2 (Goudet 2001). We regarded each city as a raccoon district group, and the adjacent districts of Kimitsu and Futtsu were combined due to the small sample size. Narita, Chiba, and Sosa were excluded from the analysis of the genetic differentiation due to the low sample size. The He among the districts were compared through a pairwise t-test using R version 3.6.2 (R Core Team 2019) and corrected via the Bonferroni method.

Individual-based genetic clustering was performed to determine the number (K) of genetically different groups (hereinafter called "cluster") using STRUCTURE 2.3 (Pritchard et al. 2000). In this analysis, the admixture model and the correlated allele frequency model were used. STRUCTURE 2.3 was run with 10 repetitions of 1,000,000 iterations of MCMC simulation, following a burn-in of 200,000 iterations at K=1–10. Using STRUCTURE Harvester (Earl & von Holdt 2012), we obtained K of the highest likelihood by considering the calculated ΔK. STRUCTURE software calculates the fractional membership of each
animal in each cluster (Q). The mean Q was calculated from 10 trials in the animal using CULUMPP 1.1.2 (Jakobsson & Rosenberg 2007) and the Q value of each animal was illustrated using district 1.1 (Rosenberg 2004). We set a criterion to include or exclude individual raccoons in particular clusters by considering the Q value of individual raccoons. We assigned individual raccoons with Q>0.7 to a single cluster; raccoons with Q<0.7 were grouped as admixtures not belonging to a particular cluster. We then calculated the district groupwise cluster frequency.

**Results**

**Mitochondrial D-loop**

Five haplotypes (A, B, C, D, and E) were identified in the Chiba Prefecture. Among them, three haplotypes (C, D, and E) were first found in Chiba and the sequences have been registered in the DNA Data Bank of Japan nucleotide sequence database (accession numbers: LC565453, LC565454, and LC565455, respectively). Only one single nucleotide polymorphism (SNP) was detected between haplotypes A and E and between haplotypes C and D. The median-joining network, including published haplotypes, is shown in Figure 2. In Noda and Narita, located in northern Chiba, haplotypes B, D, and E were identified, with haplotypes D and E being dominant. In the other district groups, haplotypes A, B, and C were identified, wherein haplotypes A and B predominated (Fig. 3A).

**Microsatellite genetic diversity among district groups**

Table 3 presents the genetic diversity index data for each district group. The value of $pA$ was highest in Noda. The values of others were similar among district groups. No significant difference was found in $He$ between district groups (Noda and Kimitsu-Futtsu, $P=0.86$; Noda and Minamiboso, $P=0.87$; all the other dyads, $P=1$).

**Comparison of the microsatellite genetic diversity with other areas**

Table 4 shows a comparison of the mean $He$ in Chiba Prefecture and that reported in previous studies. The number of loci used for comparison was different because the markers used in different studies were different. There is no significant difference between the mean $He$ between Chiba Prefecture, native areas, central Europe, Germany, and Hokkaido. Although the mean $He$ in Spain was not significantly different from that in the Chiba Prefecture, the genetic diversity tended to be lower than that in the Chiba Prefecture ($P=0.07$).

**Microsatellite genetic structure**

Based on the $\Delta K$ value, the most probable structure clustering was $K=3$. Figure 3B shows the admixture frequency and frequency of animals with Q value >0.7. Three clusters were estimated for the entire Chiba Prefecture. The Noda district group comprised Cluster I and a few admixture animals. The other district group comprised clusters II, III, and an admixture. Clusters II and III dominated the southeast and southwest areas, respectively. Figure 4 shows the fraction of the cluster (Q value) of each animal. Only
one admixture with high Cluster I proportion was detected in Noda, Narita, Sosa, and Mobara district groups. The other admixtures included clusters II and III. Table 5 shows the correspondence between the clusters and mitochondrial haplotypes. Most animals in Clusters I, II, and III had haplotypes E, A, and B, respectively.

**Discussion**

The feral raccoon, both in Japan and Europe, originated initially from animals imported from North America through pet trade (Ikeda 2004; Salgado 2018). The Invasive Alien Species Act of Japan enacted in 2005 has banned the pet trade of raccoon, and strictly regulated the transportation of living animals within Japan since 2005. Thus, the newly released or escaped captive raccoons or animals dispersed from adjacent areas should account for the origin of raccoon populations in Boso Peninsula. Boso Peninsula is almost isolated from the adjacent prefecture by the sea and the large rivers of Tone and Edo (Fig. 1); thus, either human-induced re-introduction from other areas or released or escaped captive pets should be the primary sources of feral raccoons in Boso Peninsula. Based on this unique geographic situation, we have discussed the number of possible introductions and subsequent genetic admixture processes among the locations of multiple releases or escapes by referencing the spatial genetic structure revealed in this study.

Based on the mtDNA haplotype analysis and historical records of feral raccoon, Yoshida et al. (2020) hypothesized that two independent expansions have occurred from the different founder populations, with a second expansion after the first founder had already spread over the Boso Peninsula. An alternative possibility is that, after the single release of several raccoons with different mtDNA haplotypes, genetic drift during the expansion process generated local genetic differentiation. Both processes result in spatial genetic structure after the initial expansion; however, their outcomes are different. If multiple releases with low genetic diversity within the founder animals occur at different locations, genetic admixture among the multiple founders during the subsequent expansion would eventually diminish regional genetic differentiation. Alternatively, if a single release event with high genetic diversity within the founder animals occurs in a single location, the genetic differentiation observed during the expansion process is preserved.

We detected five mtDNA haplotypes within the Chiba Prefecture, two of which have already been reported by Yoshida et al. (2020) in southern Boso Peninsula. The three newly detected haplotypes were present in tissue samples collected from the raccoons of northern Boso Peninsula. However, the SNPs between haplotypes A and E and haplotypes C and D possibly appeared after the invasion because only one SNP was identified among in each haplotype pair. This means that a third founder event occurred in addition to the first and second founder events in the southern Boso Peninsula, suggesting that at least three mitochondrial haplotypes were introduced to the Chiba Prefecture.

Three genetic groups (clusters I, II, and III) were detected in the nuclear DNA of raccoons of the Chiba prefecture using STRUCTURE analysis, and these three nuclear genetic groups corresponded to mtDNA
haplotypes E, A, and B, respectively. Therefore, these three groups, two in the east (Isumi, Cluster III and haplotype B) and west (Kimitsu-Futtsu, Cluster II and haplotype A) of southern Boso Peninsula, and one in the northwest corner (Noda, Cluster I and haplotype E) of the Chiba Prefecture (Fig. 3), are compatible with the hypothesis of at least three independent introductions within the Chiba Prefecture. Historical records of raccoon distribution and abundance in the Chiba Prefecture (Ochiai et al. 2002; Asada 2014; Yishida et al. 2020) also support the hypothesis of three independent sequential introductions and subsequent range expansion, population increase, and genetic admixture.

Even though there are significant gaps in the abundance and genetic compositions among the northwestern and southern groups (Fig. 3), the raccoons initially introduced to Isumi and Kimitsu-Futtsu were capable of steadily expanding their distribution in every direction, and genetic exchange occurred in the Tateyama-Minamiboso area, finally the genetic structure will disappear. Therefore, admitting that the currently recognizable two genetic clusters will not persist in the near future due to the absence of an effective genetic barrier between southern and northern Chiba Prefecture, such clusters can presently be treated as two different management units. Monitoring relative abundance (CPUE), sex ratio of captured raccoons, and genotypes of both the mitochondrial and nuclear DNAs enabled us to determine whether those management units were in the male dominant low density lag-phase, in the increase-phase of different genetic clusters, or in the increase-phase of the fused large cluster. In this context, we recognize that two management units in Boso Peninsula have already belonged to the increase-phase, but will soon fuse to a single genetic cluster of increase-phase to which the lag-phase population control (Asada 2013) will be difficult to apply as a feasible raccoon management practice.

Declarations

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Author Contributions Statement

This research was conceived by M. Hirose, E. Inoue and M. Hasegawa. Samples were collected by M. Hirose and K. Yoshida. M. Hirose analyzed the data and wrote the manuscript draft. All authors reviewed the manuscript.

References

1. Alda, F. et al. J. T. Genetic evidence for multiple introduction events of raccoons (Procyon lotor) in Spain. Biological Invasions, 15, 687–698(2013).

2. American Veterinary Medical Association. AVMA guidelines for the euthanasia of animals. https://www.avma.org/KB/Policies/Documents/euthanasia.pdf (2013).

3. Asada, M. “Lag-phase management” as a population management method in low density areas in sika deer (Cervus nippon) and racoon (Procyon lotor). Honyurui Kagaku (Mammalian Science). 53 (2), 243–255 (2013). (in Japanese with English Abstract).

4. Asada, M. Bayesian estimation of population size in raccoon (Procyon lotor) using state-space model based on removal sampling. Honyurui Kagaku (Mammalian Science). 54, 207–218 (2014). (in Japanese with English Abstract).

5. Assessment, M. E. Ecosystems and human well-being. United States of America: Island press. Vol. 5, p.563(2005).

6. Bagan, H. & Yamagata, Y. Landsat analysis of urban growth: How Tokyo became the world's largest megacity during the last 40 years. Remote sensing of Environment. 127, 210–222 (2012).

7. Bandelt, H. J., Forster, P. & Rohl, A. Median-joining networks for inferring intraspecific phylogenies. Molecular Biology and Evolution. 16, 37–48 (1999).

8. Biedrzycka, A., Zalewski, A., Bartoszewicz, M., Okarma, H. & Jędrzejewska, E. The genetic structure of raccoon introduced in Central Europe reflects multiple invasion pathways. Biol. Invasions. 16, 1611–1625 (2014).

9. Brownstein, M. J., Carpten, J. D. & Smith, J. R. Modulation of non-templated nucleotide addition by Taq DNA polymerase: Primer modifications that facilitate genotyping. BioTechniques. 20 (6), 1004–1010 (1996).

10. Crowl, T. A., Crist, T. O., Parmenter, R. R., Belovsky, G. & Lugo, A. E. The spread of invasive species and infectious disease as drivers of ecosystem change. Frontiers in Ecology and the Environment. 6 (5), 238–246 (2008).

11. Cullingham, C. I., Kyle, C. J. & White, B. N. Isolation, characterization and multiplex genotyping of raccoon tetranucleotide microsatellite loci. Mol. Ecol. Notes. 6 (4), 1030–1032 (2006).

12. Cullingham, C. I., Kyle, C. J., Pond, B. A. & White, B. N. Genetic structure of raccoons in eastern North America based on mtDNA: implications for subspecies designation and rabies disease dynamics. Canadian Journal of Zoology. 86, 947–958 (2008).
13. Earl, D. A. & von Holdt, B. M. STRUCTURE HARVESTER: A website and program for visualizing STRUCTURE output and implementing the Evanno method. Conservation Genetics Resources. 4 (2), 359–361 (2012).

14. Fike, J. A., Drauch, A. M., Beasley, J. C., Dharmarajan, G. & Rhodes, O. E. Development of 14 multiplexed microsatellite loci for raccoons Procyon lotor: Primer note. Mol. Ecol. Notes. 7 (3), 525–527 (2007).

15. Fischer, M. L. et al. Historical invasion records can be misleading: genetic evidence for multiple introductions of invasive raccoons (Procyon lotor) in Germany. PloS one. 10 (5), e0125441 (2015).

16. Fischer, M. L. et al. Multiple founder effects are followed by range expansion and admixture during the invasion process of the raccoon (Procyon lotor) in Europe. Diversity and Distributions. 23 (4), 409–420 (2017).

17. Gascoigne, J., Berec, L., Gregory, S. & Courchamp, F. Dangerously few liaisons: a review of mate-finding Allee effects. Popul. Ecol. 51 (3), 355–372 (2009).

18. Gehrt, S. D. Raccoons and allies. Wild mammals of North America, 611–634 (2003).

19. Goudet, J. "FSTAT, a program to estimate and test gene diversity and fixation indices (version 2.9.3)." http://www2.unil.ch/popgen/softwares/fstat.htm (2001).

20. Holleley, C. E. & Geerts, P. G. Multiplex Manager 1.0: A cross-platform computer program that plans and optimizes multiplex PCR. BioTechniques. 46 (7), 511–517 (2009).

21. Ikeda, T., Asano, M., Matoba, Y. & Abe, G. Present status of invasive alien raccoon and its impact in Japan. Global Environmental Research. 8, 125–131 (2004).

22. Jakobsson, M., Rosenberg, N. A. & CLUMPP A cluster matching and permutation program for dealing with label switching and multimodality in analysis of population structure. Bioinformatics. 23 (14), 1801–1806 https://doi.org/10.1093/bioinformatics/btm233 (2007).

23. Koressaar, T. & Remm, M. Enhancements and modifications of primer design program Primer3. Bioinformatics. 23 (10), 1289–1291 (2007).

24. Kumar, S., Stecher, G., Li, M., Knyaz, C. & Tamura, K. MEGA X: molecular evolutionary genetics analysis across computing platforms. Molecular biology and evolution. 35 (6), 1547–1549 (2018).

25. Lawson Handley, L. J. et al. Ecological genetics of invasive alien species. BioControl. 56, 409–428 (2011).

26. Lotze, J. & Anderson, S. Procyon lotor. Mammalian Species. 119, 1–8 (1979).

27. Long, J. Introduced mammals of the world: their history distribution and influence (CSIRO Publishing, 2003).

28. Mazza, G., Tricarico, E., Genovesi, P. & Gherardi, F. Biological invaders are threats to human health: an overview. Ethol Ecol Evol. 26, 112–129 (2014).

29. Ministry of the Environment Government of Japan. The birds and beasts (bears) need care habitation distribution survey in 2017. Survey report, Raccoon, Palm civet, Nutria (2018). (in Japanese)
30. Ochiai, K., Ishii, M. & Furukawa, T. Invasion and distribution of the raccoon, Procyon lotor, in Chiba Prefecture, Central Japan. Journal of the Natural History Museum and Institute, Chiba, 7, 21–27 (2002). (in Japanese)

31. Okuyama, M. W. et al. Genetic population structure of invasive raccoons (Procyon lotor) in Hokkaido, Japan: Unique phenomenon caused by pet escape or abandonment. Scientific reports. 10 (1), 1–10 (2020).

32. Parker, I. M. et al. Impact: toward a framework for understanding the ecological effects of invaders. Biological invasions. 1 (1), 3–19 (1999).

33. Peakall, R. O. D. & Smouse, P. E. GENALEX 6: genetic analysis in Excel. Population genetic software for teaching and research. Molecular ecology notes. 6 (1), 288–295 (2006).

34. Peakall, R. & Smouse, P. E. GenALEX 6.5: Genetic analysis in Excel. Population genetic software for teaching and research-an update. Bioinformatics. 28 (19), 2537–2539 (2012).

35. Pimentel, D., Zuniga, R. & Morrison, D. Update on the environmental and economic costs associated with alien invasive species in the United States. Ecol Econ. 52, 273–288 (2005).

36. Pompanon, F., Bonin, A., Bellemain, E. & Taberlet, P. Genotyping errors: causes, consequences and solutions. Nature Reviews. Genetics. 6 (11), 847–859 (2005).

37. Porter, W. F., Mathews, N. E., Underwood, H. B., Sage, R. W. & Behrend, D. F. Social organization in deer: implications for localized management. Environ. Manage. 15 (6), 809–814 (1991).

38. Pritchard, J. K., Stephens, M. & Donnelly, P. Inference of population structure using multilocus genotype data. Genetics. 155 (2), 945–959 (2000).

39. Rosenberg, N. A. Distruct: a Program for the Graphical Display of Population Structure. Mol. Ecol. Notes. 4 (1), 137–138 (2004).

40. Salgado, I. Is the raccoon (Procyon lotor) out of control in Europe? Biodiversity and conservation, 27, 2243–2256 (2018).

41. Santonastaso, T. T., Dubach, J., Hauver, S., Graser, W. H. & Gehrt, S. D. Microsatellite analysis of raccoon (Procyon lotor) population structure across an extensive metropolitan landscape. Journal of Mammalogy. 93 (2), 447–455 (2012).

42. Siripunkaw, C. et al. Isolation and characterization of polymorphic microsatellite loci in the raccoon (Procyon lotor). Molecular Ecology Resources. 8 (1), 199–201 (2008).

43. Sugai, T., Matsushima, H. & Mizuno, K. (eds) Last 400 ka landform evolution of the Kanto Plain under the influence of concurrent glacio-eustatic sea level changes and tectonic activity. Journal of Geography-Chigaku Zasshi, 122(6), 921–948 (2013). (in Japanese with English Abstract)

44. Untergasser, A. et al. Primer3-new capabilities and interfaces. Nucleic Acids Res. 40 (15), 1–12 (2012).

45. Yoshida, K., Hirose, M., Hasegawa, M. & Inoue, E. Mitochondrial DNA analyses of invasive raccoons (Procyon lotor) in the Boso Peninsula, Japan. Mammal Study. 45 (1), 1–6 (2020).
46. Vilà, M. et al. How well do we understand the impacts of alien species on ecosystem services? A pan-European, cross-taxon assessment. *Frontiers in Ecology and the Environment.* 8(3), 135–144 (2010).

47. Vilà, M. et al. How well do we understand the impacts of alien species on ecosystem services? A pan-European, cross-taxon assessment. *Frontiers in Ecology and the Environment,* 8(3), 135–144 (2010).

**Tables**

**Table 1.** Primer of the microsatellite region that we designed newly.

| Locus  | Primer sequence | Motif sequence | Accession number |
|--------|------------------|----------------|------------------|
| PLO-M3 | F-GAATGAGTCCATTTTGCTGGT | (ATCT)\textsubscript{15} | DQ388435 |
|        | R-CAGAACAGTGGGTGGAGAT |                        |                  |
| PLO-M20| F-GATTCTTATGTCTTTGGGA | (TCTA)\textsubscript{17} | DQ388437 |
|        | R-AAGTGCTTCAAGAGAAAGTGCA |                    |                  |
| PLO-M17| F-CAAGGGAGAGGAAGAACAG | (GTTT)\textsubscript{3} | DQ388440 |
|        | R-CCCCTCCCCTGTACATATTC | (TATC)\textsubscript{12} |                  |

**Table 2.** Characteristic of the microsatellite 24 locus and primer sets of multiplex PCR.
| Primer set | Marker   | N   | Range   | Na  | Ho  | He  | Label | Concentration (μM) |
|------------|----------|-----|---------|-----|-----|-----|-------|-------------------|
| Multiplex 1| PLO-M03¹ | 179 | 122-142 | 7   | 0.536 | 0.545 | FAM   | 1                 |
|            | PLO-M20¹ | 177 | 173-205 | 11  | 0.836 | 0.754 | FAM   | 2                 |
|            | PLO-M17¹ | 179 | 236-256 | 8   | 0.827 | 0.810 | FAM   | 2                 |
|            | PLO-M02¹ | 177 | 285-318 | 10  | 0.785 | 0.800 | FAM   | 1.5               |
|            | PLO-M15¹ | 179 | 109-143 | 12  | 0.832 | 0.870 | HEX   | 1                 |
|            | PLO-2-117¹ | 163 | 297-339 | 14  | 0.840 | 0.850 | HEX   | 8                 |
|            | PLO2-14¹ | 177 | 230-256 | 16  | 0.842 | 0.890 | NED   | 4                 |
| Multiplex 2| PLM17³  | 178 | 96-110  | 10  | 0.815 | 0.836 | FAM   | 2                 |
|            | PLM07³  | 179 | 155-169 | 8   | 0.771 | 0.820 | FAM   | 2                 |
|            | PLOT-11² | 178 | 201-217 | 8   | 0.708 | 0.725 | FAM   | 2                 |
|            | PLOT-08² | 178 | 244-260 | 5   | 0.624 | 0.611 | FAM   | 2                 |
|            | PLOT-03² | 176 | 302-312 | 6   | 0.699 | 0.740 | FAM   | 2                 |
|            | PLOT-04² | 164 | 344-372 | 12  | 0.616 | 0.849 | FAM   | 2                 |
|            | PLM05³  | 179 | 98-118  | 7   | 0.626 | 0.667 | HEX   | 2                 |
|            | PLM14³  | 179 | 154-168 | 8   | 0.726 | 0.761 | HEX   | 2                 |
|            | PLM08³  | 178 | 210-222 | 7   | 0.624 | 0.669 | HEX   | 4                 |
|            | PLOT-05² | 179 | 118-132 | 8   | 0.620 | 0.711 | NED   | 2                 |
| Multiplex 3| PLM09³  | 178 | 105-135 | 11  | 0.758 | 0.795 | FAM   | 2                 |
|            | PLOT-06² | 179 | 161-181 | 6   | 0.665 | 0.693 | FAM   | 2                 |
|            | PLOT-07² | 174 | 214-222 | 7   | 0.454 | 0.530 | FAM   | 2                 |
|            | PLM03³  | 179 | 130-140 | 5   | 0.547 | 0.577 | HEX   | 2                 |
|            | PLOT-02² | 179 | 184-212 | 11  | 0.799 | 0.829 | HEX   | 2                 |
|            | PLM06³  | 177 | 98-110  | 7   | 0.785 | 0.812 | NED   | 2                 |
|            | PLOT-10² | 179 | 158-182 | 7   | 0.492 | 0.565 | NED   | 2                 |
Table 3. Genetic diversity of each district.

| District       | N       | Na     | Ne     | AR   | Ho    | He    | pA  |
|----------------|---------|--------|--------|------|-------|-------|-----|
| Noda           | 13.58   | 6.58   | 4.34   | 5.18 | 0.74  | 0.75  | 22  |
| Oamishirasato  | 19.96   | 6.33   | 3.91   | 4.54 | 0.70  | 0.69  | 4   |
| Mobara         | 19.75   | 6.46   | 3.97   | 4.66 | 0.72  | 0.70  | 3   |
| Isumi          | 19.88   | 5.88   | 3.84   | 4.44 | 0.70  | 0.70  | 2   |
| Kisarazu       | 35.79   | 6.25   | 3.96   | 4.42 | 0.69  | 0.71  | 4   |
| Kimitsu-Futtsu | 21.67   | 5.50   | 3.46   | 4.17 | 0.69  | 0.67  | 1   |
| Minamiboso     | 19.71   | 5.75   | 3.55   | 4.30 | 0.67  | 0.67  | 0   |
| Tateyama       | 20.54   | 5.75   | 3.81   | 4.44 | 0.72  | 0.71  | 0   |

Table 4. The comparison results of the genetic diversity between Chiba and other areas.

$N$ = Number of the samples, Range = Range of the alleles, $N_a$ = Number of the alleles, $Ho$ = Observed heterozygosity, $He$ = Expected heterozygosity, Label = Primer Dye, Concentration ($\mu$M) = Concentration of the primer in multiplex PCR

1: Cullingham et al. (2006), 2: Fike et al. (2007), 3: Siripunkaw et al. (2008)

$N$ = Mean of the number, $Na$ = Estimated number of the alleles, $Ne$ = Effective number of alleles, $AR$ = Mean of the allelic richness, $Ho$ = Observed heterozygosity, $He$ = Expected heterozygosity, $pA$ = Number of the private alleles
| Area                        | Number of loci | $He$ | $He$ of Chiba | $P$-value | Theses               |
|-----------------------------|----------------|------|---------------|------------|----------------------|
| Native area                 |                |      |               |            |                      |
| Ontario                     | 7              | 0.84 | 0.79          | 1.00       | Cullingham et al. 2006 |
| Indiana                     | 9              | 0.80 | 0.70          | 0.18       | Fike et al. 2007     |
| Columbia and Missouri       | 8              | 0.84 | 0.74          | 0.29       | Siripunkaw et al. 2008 |
| Illinois                    | 7              | 0.83 | 0.79          | 1.00       | Santonastaso et al. 2012 |
| Missouri                    | 8              | 0.81 | 0.74          | 0.29       | Alda et al. 2013     |
| Invasion area               |                |      |               |            |                      |
| Germany, Czech, and Poland  | 11             | 0.66 | 0.71          | 0.23       | Biedrzycka et al. 2013 |
| Germany                     | 16             | 0.70 | 0.72          | 0.21       | Fischer et al. 2015  |
| Spain                       | 8              | 0.62 | 0.74          | 0.07       | Alda et al. 2013     |
| Hokkaido                    | 5              | 0.78 | 0.76          | 1.00       | Okuyama et al. 2020  |

Number of loci = The number of loci that we used for comparison, $He$ = Mean of expected heterozygosity, $P$-value = The P-value that was calculated as a result of the sign test

Table 5. The correspondence of cluster and haplotypes. The numerals are the number of individuals.

| Cluster | Haplotype | Total |
|---------|-----------|-------|
|         | A | B | C | D | E |     |
| I       | 0 | 1 | 0 | 4 | 8 | 13  |
| II      | 36| 31| 0 | 0 | 0 | 67  |
| III     | 25| 49| 0 | 0 | 0 | 74  |
| Admixture | 9 | 13| 1 | 2 | 9 | 25  |
| Total   | 70| 94| 1 | 6 | 8 | 179 |

**Figures**
Figure 1

Around the study area. Partially modified based on the map of Geospatial Information Authority of Japan. Note: The designations employed and the presentation of the material on this map do not imply the expression of any opinion whatsoever on the part of Research Square concerning the legal status of any country, territory, city or area or of its authorities, or concerning the delimitation of its frontiers or boundaries. This map has been provided by the authors.
Figure 2

The median-joining network of mitochondrial DNA sequences of raccoons. Haplotypes A, B, C, D, and E were detected in Chiba prefecture. White circles represent haplotypes in the study by Cullingham et al. (2008). We followed the categorization (lineages I, II, and III) of Cullingham et al. (2008).
Figure 3

The background maps of (A) and (B) show the distribution of raccoons in Chiba prefecture in 2017. The maps are based on the CPUE (captures per 100 trap days) value in each administrative district (Chiba prefecture unpublished). (A) The distribution and the frequency of haplotypes observed in Chiba prefecture. (B) The frequency and the distribution of clusters (K=3) in each district groups. Note: The designations employed and the presentation of the material on this map do not imply the expression of
any opinion whatsoever on the part of Research Square concerning the legal status of any country, territory, city or area or of its authorities, or concerning the delimitation of its frontiers or boundaries. This map has been provided by the authors.

![Map diagram](image)

**Figure 4**

The fraction of the cluster was estimated by STRUCTURE (K=3). The individuals are expressed in one bar. The color of cluster 1 is black, cluster 2 is grey and cluster 3 is white. The length of color bars expresses a fraction of the cluster (Q value).