Husk of Agarwood Fruit-Based Hydrogel Beads for Adsorption of Cationic and Anionic Dyes in Aqueous Solutions

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Abstract: Hydrogel beads based on the husk of agarwood fruit (HAF)/sodium alginate (SA), and based on the HAF/chitosan (CS) were developed for the removal of the dyes, crystal violet (CV) and reactive blue 4 (RB4), in aqueous solutions, respectively. The effects of the initial pH (2–10) of the dye solution, the adsorbent dosage (0.5–3.5 g/L), and contact time (0–540 min) were investigated in a batch system. The dynamic adsorption behavior of CV and RB4 can be represented well by the pseudo-second-order model and pseudo-first-order model, respectively. In addition, the adsorption isotherm data can be explained by the Langmuir isotherm model. Both hydrogel beads have acceptable adsorption selectivity and reusability for the study of selective adsorption and regeneration. Based on the effectiveness, selectivity, and reusability of these hydrogel beads, they can be treated as potential adsorbents for the removal of dyes in aqueous solutions.

Keywords: agarwood; hydrogel; adsorption; reactive blue 4; crystal violet

1. Introduction

With the rapid development of industry and population growth, natural resources cannot meet the demand of human life. The extensive use of synthetic materials and the discharge of wastewater from manufacturing industries have seriously polluted water resources and the environment. Industrial wastewater contains organic compounds and heavy metals, which are water pollutants that cause worldwide environmental problems [1]. Synthetic dyes are common organic pollutants in industrial wastewater. Most of these dyes have diverse structures and stable properties and are not easily biodegraded [2]. The effective removal of dyes from wastewater is one of the major environmental challenges in water management since many dyes are carcinogenic, mutagenic, allergenic, and highly toxic to aquatic living organisms [3]. The methods developed for treating pollutants in wastewater currently include adsorption, membrane separation, reverse osmosis, photodegradation, ozone treatment, oxidation treatment, biological treatment, coagulation and flocculation, chemical conversion, electrochemical methods, etc. [4]. The adsorption method for wastewater treatment has the advantages of inexpensiveness, high efficiency, easy operating procedures, system flexibility, wide-range applicability, insensitivity to pollutants, and facilitation of material recovery [5,6]. Various low-cost, environmentally friendly alternative adsorbents have been prepared from agricultural waste or byproducts for the removal of dyes from wastewater, such as mung bean husk [7], corn stalk [8], rice bran [9], bamboo [10], rice husk [11], and Artocarpus odoratissimus leaves [12]. These green adsorbents have the major characteristic properties of low cost with satisfactory adsorption capacities and environmentally friendly nature (such as derived from abundant natural sources, biodegradable, non-toxic, etc.), which have the lowest impact on the environmental ecosystem [13].
Recently, hydrogel-based adsorbents synthesized from bioresources such as chitosan, sodium alginate, and starch cellulosic materials have attracted particular attention for dye removal because of their three-dimensional network structure, high availability, environmental friendliness, high swelling property, versatility as a treatment medium, and simple recoverability [14–16]. Sodium alginate (SA) is a natural polysaccharide composed of β-D-mannuronate units and α-L-guluronate units and can be extracted from seaweed [17]. Chitosan (CS), a copolymer polysaccharide of deacetylated units and acetylated units, is prepared mainly by deacetylation of chitin [18]. SA has a high content of carboxyl groups, and CS has a large number of amino (-NH$_2$) and hydroxyl (-OH) groups that can be chemically or physically reacted with a crosslinking agent to prepare hydrogels [19,20]. In recent years, many studies have pointed out that hydrogels based on SA and CS have good adsorption capacities for pollutants such as metal ions, antibiotics, and organic pollutants in aqueous media and are efficient for wastewater treatment [21–24]. However, the low mechanical strength of hydrogels limits its commercial applications as adsorbents. Agarwood is the most expensive wood in the world and is widely planted in Southeast Asia [25]. Agarwood fruit has a pear-like shape and a hollow structure inside. The husk of agarwood fruit (HAF) is quite hard and generally discarded as agricultural waste due to good mechanical strength, superior adsorption potential, and availability [26]. The introduction of HAF into chitosan may improve the swelling strength and stability of the adsorbent. Crystal violet (CV) dye is a tri-phenylmethane dye and reactive blue 4 (RB4) dye is an anthraquinone-based chlorotriazine dye. Both of them are often used in the textile industry and are known for their carcinogenic, mutagenic, and mitotic poisoning nature [27,28]. Therefore, this work attempts to prepare hydrogel beads based on the husk of agarwood fruit, SA, and CS for the removal of CV and RB4 dyes in aqueous solutions. In addition, the kinetics, isotherms, selectivity, and regeneration ability of the hydrogel beads were investigated to evaluate their adsorption properties.

2. Results and Discussion
2.1. Characterization

Scanning electron microscopy (SEM) was applied to study the morphologies of the hydrogel beads. Figure 1 shows the SEM images of hydrogel beads. It can be seen that the particle size of all the hydrogel beads is approximately 1.5 mm. Before chemical modification, the surface of HAF-SA is smooth, and that of HAF-CS is rough (Figure 2a,b). After modification, the surface of HAF-SA becomes wrinkled, and that of HAF-CS becomes smooth (Figure 2c,d). In addition, more clearly spherical shapes could be observed for the MHAF-CS hydrogel beads. From the high-magnification SEM images, the wrinkled surface of hydrogel beads was observed due to the incorporation of HAF into the hydrogel beads. The functional groups in HAF interacted strongly with SA and CS and resulted in a higher crosslinking degree [29].

FTIR analysis was conducted to identify the functional groups on the surface of the hydrogel beads. The FTIR spectra of the MHAF-SA hydrogel beads before and after CV adsorption are shown in Figure 1a. In the spectrum, the broad peak of MHAF-SA at 3370 cm$^{-1}$ was attributed to the stretching vibrations of the –OH groups from pure SA and HAF molecules. The peaks at 1610 cm$^{-1}$, 1410, and 1028 cm$^{-1}$ are assigned to the –COO$^-$ (asymmetric), –COO$^-$ (symmetric), and C-O-C stretching vibrations, respectively [30]. After modification by acetic acid, the intensity of these peaks in the spectrum of the MHAF-SA hydrogel beads was increased compared with that of the HAF-SA hydrogel beads. In addition, new peaks were generated at 1718 cm$^{-1}$ (C=O) and 1238 cm$^{-1}$ (C-O), indicating more incorporation of –COO$^-$ groups. All the peaks associated with –COO$^-$ shifted in the spectra of MHAF-SA after CV adsorption and indicated electrostatic interactions between the negatively charged carboxyl groups of MHAF-SA and the positively charged CV dye. Figure 1b displays the FTIR spectra of MHAF-CS hydrogel beads before and after RB4 adsorption. The broad peak of MHAF-CS at approximately 3350 cm$^{-1}$ can be assigned to the overlapping OH and NH stretching vibrations contributed by pure CS and...
HAF. Other characteristic peaks at 2872, 1590, 1373, and 1027 cm$^{-1}$ are attributed to C-H stretching, N-H bending, C-H bending, and C-O-C stretching vibrations, respectively. After modification by sodium hydroxide, the peaks at 3350, 1590, and 1027 cm$^{-1}$ were obviously enhanced compared with those of HAF-CS. After RB4 adsorption, the peaks at 3350, 1590, and 1027 cm$^{-1}$ shifted to 3372, 1567, and 1016 cm$^{-1}$, respectively. In addition, a new peak appears at 1160 cm$^{-1}$ after RB4 adsorption, which corresponds to the $-\text{SO}_3^-$ groups of RB4 molecules [31]. According to this information, RB4 adsorption onto MHAFF-CS hydrogel beads may be dependent on $-\text{OH}$ and $-\text{NH}_2$ groups.

Figure 3 shows the point of zero charge (pH$_{pzc}$) of the MHAFF-SA and MHAFF-CS hydrogel beads. The pH$_{pzc}$ values of the MHAFF-SA and MHAFF-CS hydrogel beads were 2.50 and 7.35, respectively. Therefore, the surface of MHAFF-CS was more positive than MHAFF-SA due to the presence of $-\text{NH}_2$ groups, as proved by the FTIR spectra. MHAFF-SA can be used as a cation attractor in solutions with pH values above the pH$_{pzc}$ and MHAFF-CS can be treated as an anion attractor in solutions with pH values below the pH$_{pzc}$ [32]. The swelling behavior of hydrogel beads was studied, and the results are shown in Figure 4. As displayed in Figure 4, compared to CS, sodium alginate (SA) exhibits a much lower swelling ratio (5\%) and does not change much with increasing time. The surface of the SA hydrogel bead is hydrophobic, and water molecules do not easily enter the interior. Thus, the swelling ratio of SA is low and does not change obviously (Figure 4a). The swelling ratio of MHAFF-SA increased significantly with the addition of HAF due to certain hydrophilic groups in HAF allowing water molecules to enter the beads. The higher swelling property can improve dye adsorption since more water molecules can penetrate into the beads [33,34]. The chitosan structure can generate hydrogen bonds with water molecules. Therefore, the chitosan (CS) hydrogel beads have good water adsorption capacity. The swelling ratio of MHAFF-CS decreased with the addition of HAF, which is less hydrophilic than chitosan (Figure 4b). Although the swelling ratio of MHAFF-CS is lower than that of chitosan, it was better for adsorption performance since the introduction of HAF into chitosan improved the swelling strength and stability [34].

2.2. Effect of pH

The pH of the solution is one of the most important influencing factors in the adsorption system and can change the surface charge of the adsorbent. The pH effect on the dye adsorption of hydrogel beads at a fixed dosage of 30 mg and an initial dye concentration of 200 mg/L is shown in Figure 5. The adsorption capacity of CV adsorbed onto MHAFF-SA increased as the pH increased, while RB4 adsorbed onto MHAFF-CS had the opposite effect. At pH values higher than 2.5 (pH$_{pzc}$ of MHAFF-SA), the surface of the hydrogel bead is negatively charged and prefers to adsorb the cationic dye (CV). The surface of the adsorbent is occupied by a large amount of H$^+$ ions under acidic conditions, which compete for the adsorption sites with CV and result in a lower adsorption capacity. Moreover, the cationic dye easily reacts with oxygenic functional groups ($-\text{COO}^-$ and $-\text{OH}^-$ groups) in a high pH dye solution [30]. At pH values lower than 7.35 (pH$_{pzc}$ of MHAFF-CS), the surface of the hydrogel bead is positively charged due to the protonation of amino groups ($-\text{NH}_3^+$) and facilitates RB4 ($\text{SO}_3^-$ groups) adsorption by electrostatic attraction. MHAFF-CS tends to dissolve at pH 2, which reduces the adsorption capacity of RB4. In addition, the precipitation of CV was observed at pH 10. The optimum pH values for the adsorption of CV and RB4 were determined to be 9 and 3 for MHAFF-SA and MHAFF-CS, respectively.
Figure 1. FTIR spectra of (a) MHAF-SA hydrogel beads and (b) MHAF-CS hydrogel beads.
Figure 2. SEM images of hydrogel beads: (a) HAF-SA, (b) HAF-CS, (c) MHAF-SA, and (d) MHAF-CS.
Figure 3. The $pH_{pzc}$ of (a) MHAF-SA hydrogel beads and (b) MHAF-CS hydrogel beads.
2.2. Effect of pH

The pH of the solution is one of the most important influencing factors in the adsorption system and can change the surface charge of the adsorbent. The pH effect on the dye adsorption of hydrogel beads at a fixed dosage of 30 mg and an initial dye concentration of 100 mg/L is shown in Figure 5. As shown in Figure 5, the removal efficiency of CV and RB4 increased with increasing pH, while the adsorption capacity was the opposite. Therefore, the optimum pH for both hydrogel beads to adsorb dyes was 7.

Figure 4. The swelling behaviors of (a) MHAF-SA hydrogel beads and (b) MHAF-CS hydrogel beads.

2.3. Effect of Dosage

The dosage of the adsorbent is another important key factor that affects the adsorption system to reduce the cost of wastewater treatment and maximize the benefit of the adsorbent. As shown in Figure 6, the removal efficiency (RE %) of both CV and RB4 increased with the increasing dosage, while the adsorption capacity was the opposite. When the dosage of hydrogel beads was 2.5 g/L, the removal efficiencies of CV and RB4 increased insignificantly with increasing dose. Therefore, the optimum dosage for both hydrogel beads to adsorb dyes was 2.5 g/L.
The concentration of 200 mg/L is shown in Figure 5. The adsorption capacity of CV adsorbed onto MHAF-SA increased as the pH increased, while RB4 adsorbed onto MHAF-CS had the opposite effect. At pH values higher than 2.5 (pH pzc of MHAF-SA), the surface of the hydrogel bead is negatively charged and prefers to adsorb the cationic dye (CV). The surface of the adsorbent is occupied by a large amount of H+ ions under acidic conditions, which compete for the adsorption sites with CV and result in a lower adsorption capacity. Moreover, the cationic dye easily reacts with oxygenic functional groups (–COO− and –OH− groups) in a high pH dye solution [30]. At pH values lower than 7.35 (pHpzc of MHAF-CS), the surface of the hydrogel bead is positively charged due to the protonation of amino groups (–NH3+) and facilitates RB4 (SO3− groups) adsorption by electrostatic attraction. MHAF-CS tends to dissolve at pH 2, which reduces the adsorption capacity of RB4. In addition, the precipitation of CV was observed at pH 10. The optimum pH values for the adsorption of CV and RB4 were determined to be 9 and 3 for MHAF-SA and MHAF-CS, respectively.

Figure 5. Effect of pH on dye adsorption.

The dosage of the adsorbent is another important key factor that affects the adsorption system to reduce the cost of wastewater treatment and maximize the benefit of the adsorbent. As shown in Figure 6, the removal efficiency (RE %) of both CV and RB4 increased with the increasing dosage, while the adsorption capacity was the opposite. When the dosage of hydrogel beads was 2.5 g/L, the removal efficiencies of CV and RB4 increased insignificantly with increasing dose. Therefore, the optimum dosage for both hydrogel beads to adsorb dyes was 2.5 g/L.

Figure 6. Effect of adsorbent dosage on the removal of (a) CV onto MHAF-SA and (b) RB4 onto MHAF-CS.

The effect of contact time on the adsorption of CV and RB4 onto hydrogel beads was investigated with a dosage of 2.5 g/L and an initial dye concentration of 200 mg/L at optimum pH values and 40 °C. Figure 7 shows the influence of the contact time on the dye adsorption capacities of MHAF-SA and MHAF-CS. Both hydrogel beads have similar trends for adsorption capacity, which increases with time. The adsorption, in the beginning, was very fast since a large number of available active sites or free functional groups can attract dye molecules, resulting in the adsorption capacities increasing significantly. Subsequently, the adsorption of CV and RB4 onto MHAF-SA and MHAF-CS became slower and tended to be stable, respectively, after 360 min and 420 min, finally approaching an equilibrium state. The slow adsorption can be ascribed to the active sites or free functional groups gradually being filled and the accumulation of dye molecules in the hydrogel beads, which makes dye adsorption onto the hydrogel beads more difficult.
2.4. Effect of Contact Time

The effect of contact time on the adsorption of CV and RB4 onto hydrogel beads was investigated with a dosage of 2.5 g/L and an initial dye concentration of 200 mg/L at optimum pH values and 40 °C. Figure 7 shows the influence of the contact time on the dye adsorption capacities of MHAF-SA and MHAF-CS. Both hydrogel beads have similar trends for adsorption capacity, which increases with time. The adsorption, in the beginning, was very fast since a large number of available active sites or free functional groups can attract dye molecules, resulting in the adsorption capacities increasing significantly. Subsequently, the adsorption of CV and RB4 onto MHAF-SA and MHAF-CS became slower and tended to be stable, respectively, after 360 min and 420 min, finally approaching an equilibrium state. The slow adsorption can be ascribed to the active sites or free functional groups gradually being filled and the accumulation of dye molecules in the hydrogel beads, which makes dye adsorption onto the hydrogel beads more difficult.

![Figure 7. Effect of contact time on dye adsorption.](image)

2.5. Adsorption Kinetics

To study the effectiveness of the hydrogel beads, the adsorption kinetics of CV and RB4 onto the hydrogel beads were analyzed with pseudo-first-order model (PFOM, expressed by Equation (1)) and pseudo-second-order model (PSOM, expressed by Equation (2)) [35,36].

\[
\ln (q_e - q_t) = \ln q_e - k_1 t
\]

\[
\frac{t}{q_t} = \frac{1}{k_2 q_e^2} + \frac{t}{q_e}
\]

where \(k_1\) (1/min) and \(k_2\) (g/mg·min) are the adsorption rate constants of pseudo-first-order and pseudo-second-order models, respectively; \(q_t\) (mg/g) and \(q_e\) (mg/g) are the adsorption capacities at time \(t\) (min) and adsorption equilibrium, respectively.

The parameters of the kinetic models and correlation coefficients are summarized in Table 1. As shown in Figure 8, the pseudo-second-order model is more suitable than the pseudo-first-order model to describe the kinetic data of CV onto MHAF-SA hydrogel beads. However, compared to the pseudo-second-order model, the pseudo-first-order model yields a better fit for RB4 onto MHAF-CS. The results indicated that the kinetic data of CV adsorption onto MHAF-SA agreed well with the calculated values by the pseudo-second-order model, and the adsorption behavior for RB4 onto MHAF-CS obeyed
the pseudo-first-order model. The pseudo-first-order model describes the adsorption in solid-liquid systems based on the sorption capacity of solids and the pseudo-second-order kinetics model represents chemisorption that involved chemical reactions between adsorbent and adsorbate [37].

Table 1. Kinetic parameters for the adsorption of dyes onto hydrogel beads at 40°C.

| Dye | CV       | RB4       |
|-----|----------|-----------|
|     | \(q_{exp}\) (mg/g) |         |
| PFOM| 97.56 | 80.41 |
|     | \(k_1\) (1/min) |         |
|     | 0.0068 | 0.0042 |
|     | \(q_e\) (mg/g) |         |
|     | 63.7 | 83.0 |
|     | \(R^2\) |         |
|     | 0.8353 | 0.9739 |
| PSOM| \(k_2\) (g/mg·g) |         |
|     | \(1.92 \times 10^{-4}\) | \(1.20 \times 10^{-4}\) |
|     | \(q_e\) (mg/g) |         |
|     | 104.17 | 84.03 |
|     | \(R^2\) |         |
|     | 0.9715 | 0.9635 |

Figure 8. Kinetics models for dyes adsorption: (a) pseudo-first-order model, (b) pseudo-second-order model.
2.6. Adsorption Isotherms

The adsorption isotherm study provides details about the interaction between the dye molecules and hydrogel beads at equilibrium. Various initial CV and RB4 dye concentrations (20–600 mg/L) at constant temperature were evaluated. The linear forms of isotherm models such as the Langmuir (expressed by Equation (3)) [38] and Freundlich (expressed by Equation (4)) [39] models were used to fit the adsorption isotherm data.

\[
\frac{C_e}{q_e} = \frac{1}{q_m K_L} + \frac{C_e}{q_m} \quad \text{(3)}
\]

\[
\ln q_e = \frac{1}{n} \ln C_e + \ln K_F \quad \text{(4)}
\]

where \(C_e\) (mg/L) is the dye equilibrium concentration in aqueous solution; \(q_m\) (mg/g) and \(q_e\) (mg/g) are the maximum adsorption capacity and adsorption equilibrium, respectively; \(K_L\) (L/mg) is the Langmuir constant; \(K_F\) and \(n\) are the Freundlich isotherm constants and adsorption intensity, respectively. The plots are illustrated in Figure 9, and the corresponding isotherm parameters are shown in Table 2.

Based on the linear regression coefficient (\(R^2\)), the adsorption of CV onto MHAF-SA and RB4 onto MHAF-CS correlated well with the Langmuir isotherm model, indicating that the dyes CV and RB4 were mainly adsorbed onto a homogeneous surface of hydrogel beads and had monolayer coverage with uniform adsorption energies. The maximum monolayer adsorption capacity (\(q_m\)) values from the Langmuir isotherm model for CV adsorbed onto MHAF-SA and RB4 adsorbed onto MHAF-CS were compared with other adsorbents reported in the literature and presented in Table 3. As can be seen from this table, the MHAF-SA and MHAF-CS showed higher adsorption capacity for CV and RB4, suggesting that these adsorbents can be applied potentially in the adsorption system.

Table 2. Isotherm parameters for dye adsorption onto hydrogel beads.

| Dye  | Temperature (°C) | CV     | RB4    |
|------|------------------|--------|--------|
|      | 30               | 40     | 50     |
|      | Langmuir         |        |        |
|      | \(q_m\) (mg/g)   | 232.56 | 312.50 | 370.37 |
|      | \(K_L\) (L/mg)   | 0.022  | 0.018  | 0.016  |
|      | \(R^2\)          | 0.975  | 0.989  | 0.982  |
|      | Freundlich       |        |        |
|      | \(K_F\) (mg/g)(L/mg)^1/n | 22.43 | 19.60  | 16.39  |
|      | \(n\)            | 2.41   | 1.98   | 1.74   |
|      | \(R^2\)          | 0.875  | 0.936  | 0.957  |

Table 3. Comparison of adsorption capacity of various adsorbents.

| Dye | Adsorbent                               | Adsorption Capacity (mg/g) | Reusability (Five Cycle) | Reference |
|-----|-----------------------------------------|---------------------------|--------------------------|-----------|
| CV  | NaOH-modified rice husk                 | 44.87                     | -                        | [40]      |
|     | Chitosan magnetic composite microspheres| 86.6                      | -1.6%                    | [41]      |
|     | Acid-activated sintering process red mud| 336.3                     | -                        | [42]      |
|     | Chitosan pyrrole                        | 150.16                    | -                        | [43]      |
|     | Cellulose-based adsorbent               | 182.15                    | >−25%                    | [44]      |
|     | MHAF-SA                                 | 232.56–370.37             | −1.69%                   | This study|
| RB4 | Heat-treated fungal biomass             | 156.9                     |                          | [45]      |
|     | Bokbunja seed wastes                    | 26.1                      |                          | [46]      |
|     | rice bran/ Fe₃O₄                         | 185.19                    | −4.6%                    | [9]       |
|     | Activated carbon                        | 131.9                     |                          | [47]      |
|     | rice bran/SnO₂/Fe₃O₄                    | 218.82                    | −12%                     | [48]      |
|     | MHAF-CS                                 | 156.25–270.27             | ~−20%                    | This study|
$$C_e q_e = 1 q_m K_L + C_e q_m$$ (3)

$$\ln q_e = 1 n \ln C_e + \ln K_F$$ (4)

where \(C_e\) (mg/L) is the dye equilibrium concentration in aqueous solution; \(q_m\) (mg/g) and \(q_e\) (mg/g) are the maximum adsorption capacity and adsorption equilibrium, respectively; \(K_L\) (L/mg) is the Langmuir constant; \(K_F\) and \(n\) are the Freundlich isotherm constants and adsorption intensity, respectively. The plots are illustrated in Figure 9, and the corresponding isotherm parameters are shown in Table 2.

2.7. Selective Adsorption of Hydrogel Beads

In real applications, the selective removal of pollutants from wastewater is important [32]. Three groups of dye mixtures, namely, RB4/CR, CV/NR, and RB4/CV, were used to study the selectivity of the MHAF-SA and MHAF-CS hydrogel beads. As shown in Figure 10a, the absorption peaks of the mixed dyes (RB4/CR) showed no obvious change after adsorption. In addition, the peak shape was not similar to any absorption spectra of a single dye, which means that MHAF-SA has poor adsorption capacity for anionic dyes due to the electrostatic repulsion between the dyes and hydrogel beads. Figure 10b shows the absorption spectra results of MHAF-SA before and after adsorption of the cationic mixed dyes (CV/NR). After adsorption, the intensity of the absorption peaks of the mixed dyes decreased obviously, indicating that MHAF-SA has good adsorption performance for cationic dyes. As presented in Figure 10c, the absorption peak corresponding to CV (cationic dye) in the mixed dye solution disappeared after adsorption, demonstrating that...
the cationic dye could be selectively removed by MHAF-SA. Figure 10d shows the results of RB4/CR adsorbed onto MHAF-CS. After adsorption, the intensity of the mixed dyes decreased, and the absorption spectrum was not similar to that of any single dye, indicating that MHAF-CS can effectively adsorb anionic dyes containing SO$_3^-$ groups. As shown in Figure 10e, there was no obvious change in the mixed dye (CV/NR) before and after adsorption, implying that MHAF-CS does not support the adsorption of cationic dyes. Figure 10f shows the change in the absorption peak before and after the adsorption of the mixed dye (RB4/CV). After absorption, the absorption peak intensity of the mixed dye decreased, and the absorption spectra were close to those of the CV dye. Therefore, MHAF-SA and MHAF-CS have acceptable adsorption selectivity towards cationic and anionic dyes, respectively.

![Figure 10](image_url)

**Figure 10.** The absorption spectra of the dye mixture onto MHAF-SA (a–c) and MHAF-CS (d–f).
2.8. Regeneration of Hydrogel Beads

The regeneration performance of the hydrogel beads is shown in Figure 11. As presented in this figure, the CV removal efficiency of MHA-F-SA is not obviously decreased for up to five cycles, indicating that MHA-F-SA has high reusability. Although the RB4 removal efficiency of MHA-F-CS decreased by approximately 20% under the same conditions, MHA-F-CS still retained an RE (%) over 70%, suggesting the reusability of MHA-F-CS hydrogel beads. The reusability of MHA-F-SA and MHA-F-CS is also compared with the report in the literature, as shown in Table 3. The reusability of MHA-F-SA is excellent for CV removal, and there is acceptable reusability of MHA-F-CS for RB4 removal. These results indicate that the hydrogel beads prepared in this work can be regenerated and reused to efficiently and selectively remove CV and RB4 dyes.

![Figure 11. Regeneration performance of MHA-F-SA and MHA-F-CS.](image)

2.9. Adsorption Mechanism

According to the FTIR spectra of MHA-F-SA and MHA-F-CS before and after adsorption of dyes, the effect of pH discussion, and based on the surface properties of the adsorbents, the adsorption mechanism of CV onto MHA-F-SA and RB4 onto MHA-F-CS are illustrated in Figure 12. The CV adsorption onto the surface of MHA-F-SA may be due to the cationic dye easily reacting with oxygenic functional groups (–COO\(^-\) and \(\text{–OH}^-\) groups) in a high pH dye solution (see Figure 12a). Under acidic conditions, the surface of MHA-F-CS is positively charged due to the protonation of amino groups (\(\text{–NH}_3^+\)) and facilitates RB4 (SO\(_3^-\) groups) adsorption by electrostatic attraction (Figure 12b). Therefore, the adsorption of CV onto MHA-F-SA and RB4 onto MHA-F-CS might contain the electrostatic attraction and hydrogen bonding combination [40,49].

![Figure 12. Adsorption mechanism of (a) CV onto MHA-F-SA and (b) RB4 onto MHA-F-CS.](image)
3. Materials and Methods

3.1. Materials

Acetic acid (purity $\geq 99.83\%$), ethanol (purity $\geq 99.8\%$), ammonia hydroxide (purity $\geq 35\%$), and sodium hydroxide (purity $\geq 97\%$) were purchased from Fisher Chemical (Pittsburgh, PA., USA). Sodium chloride (purity $> 99.5\%$), Congo red (CR), and crystal violet (CV) were purchased from Acros Organics (Geel, Belgium). Alginic acid sodium salt powder, chitosan, reactive blue 4 (RB4), and neutral red (NR) were supplied by Sigma-Aldrich (Natick, MA., USA). Hydrochloric acid (purity $>37\%$) was obtained from Scharlau (Senmanat, Spain). Calcium chloride (purity $\geq 89\%$) was purchased from Avantor (Allentown, PA., USA). All chemicals were analytical grade and used without further purification. Agarwood (Aquilaria agallocha Roxb) fruit was obtained from the local farmer in Kaohsiung, Taiwan.

3.2. Agarwood Fruit Treatment

The treatment process of agarwood fruit is shown in Figure 13. First, the agarwood fruit was washed repeatedly with water four to five times to remove surface dirt. After cleaning, the fruit was exposed to the sun for 3–4 days to dry naturally. The husk was obtained from the fruit and was ground and sieved (<100 mesh) for adsorbent preparation.

3.3. Preparation of HAF-SA Hydrogel Beads

HAF powder (400 mg) was added to 40 mL of deionized water and stirred on a magnetic stirrer for 30 min. Then, the solution (10 mL) was mixed with 200 mg of sodium alginate powder and stirred continuously for 1 h to completely dissolve the powder. The solution was dropped into a calcium chloride solution (5 wt%) via a syringe needle at a speed of 75 mL/h. The HAF-SA hydrogel beads were formed by an ion exchange process. After washing with deionized water, the hydrogel beads were immersed in 4.4 M acetic acid solution at 30 °C for 5 h at an agitation rate of 100 rpm in a shaker with a thermostatic water bath. Then, the modified HAF-SA (MHAF-SA) hydrogel beads were filtered and washed with deionized water at least three times. Finally, the MHAF-SA hydrogel beads were obtained after drying in an oven at 50 °C for 6 h.

3.4. Preparation of HAF-CS Hydrogel Beads

The method to prepare chitosan hydrogels by physical crosslinking is preferred over the chemical crosslinking method due to the prevention of potential toxicity from chemical
crosslinking agents [50]. Therefore, physically crosslinked chitosan hydrogels were applied in this study. First, chitosan solution was prepared by dissolving 1.2 g of chitosan powder in 50 mL of (1%, v/v) acetic acid solution. The solution was stirred continuously at 40 °C for 2 h and then remained undisturbed for 12 h to release air bubbles. Next, HAF powder (400 mg) was added to the chitosan solution and stirred for 30 min. The solution was dropped into a 0.5 M sodium hydroxide solution via a syringe needle with a speed of 75 mL/h. The hydrogel beads were formed and stirred for 3 h. Subsequently, the HAF-CS hydrogel beads were filtered, washed, and immersed in 4.4 M ammonia solution at 30 °C for 5 h at an agitation rate of 100 rpm in a shaker with a thermostatic water bath. After filtration, washing, and drying at 50 °C for 6 h, modified HAF-CS (MHAF-CS) hydrogel beads were obtained.

3.5. Characterization

The morphologies of the hydrogel beads were observed by scanning electron microscopy (SEM), and the functional groups on the surface were analyzed by Fourier transform infrared (FTIR) spectroscopy. The point of zero charge (pH_{pzc}) was determined using the pH-drift method [51] to investigate the surface charge of the hydrogel beads. To test the swelling ability of the hydrogel beads, a gravimetric method was applied [52]. A certain amount of hydrogel beads was immersed in deionized water at room temperature for a period of 6 h. The swelling ratio (S%) was determined by the following equation [53]:

\[
S(\%) = \frac{W_w - W_d}{W_d} \times 100\%
\]  

where \(W_w\) and \(W_d\) are the weights (g) of wet and dried hydrogel beads, respectively.

3.6. Batch Adsorption Experiments

The dye adsorption experiments were performed on a BT-150D shaker (YIH DER Co., Ltd., Taiwan) with a shaking speed of 100 rpm via a batch adsorption process. Equilibrium adsorption experiments were carried out by adding 30 mg of hydrogel beads into 20 mL of different concentrations of dye solution at 30 °C. After the adsorption approached equilibrium, the content of dye adsorbed onto the hydrogel beads was analyzed at the maximum wavelength (RB4 = 595 nm; CV = 590 nm) by using a UV/VIS spectrophotometer (Thermo Scientific, Waltham, MA, USA). The pH effect was investigated by adjusting the initial pH of RB4 or CR solutions in the range of 2–10 while the initial dye concentration was fixed at 200 mg/L. The dosage effect of hydrogel beads was studied from 0.5 to 3.5 g/L with an initial dye concentration of 200 mg/L. The temperature effect on the adsorption was studied from 30 to 60 °C with an initial dye concentration of 200 mg/L. The isothermal adsorption experiments were conducted with 20 mL of dye solutions with different initial concentrations (200–600 mg/L) at a fixed temperature (30 °C). All adsorption experiments were carried out in triplicate to confirm the accuracy of this study. The adsorption capacity (\(q\)) and the removal efficiency (\(RE\)% of RB4 or CR adsorbed onto hydrogel beads RB4 were calculated by the following equations:

\[
q = \frac{(C_0 \times C_e)V}{W}
\]

\[
RE(\%) = \frac{C_0 \times C_e}{C_0} \times 100\%
\]

where \(C_0\) and \(C_e\) are the initial and equilibrium dye concentrations (mg/L), respectively, \(V\) (L) is the volume of the dye solution, and \(W\) (g) is the dry weight of the hydrogel beads used.

3.7. Selective Adsorption Study

Textile wastewater contains different types of dyes; thus, selective adsorption studies of adsorbents are quite important for practical applications. Three groups of dye mix-
tures, namely, RB4/CR (anionic dye-anionic dye), CV/NR (cationic dye-cationic dye), and RB4/CV (anionic dye-cationic dye), were used to study the adsorption selectivity of MHAF-SA and MHAF-CS hydrogel beads. The optimum dosage of hydrogel beads (2.5 g/L) was added to 20 mL of dye solution (200 mg/L) at a fixed temperature. After the adsorption reached equilibrium, the residual concentration of the dye mixture was monitored by a UV/VIS spectrophotometer.

3.8. Regeneration Study

To reduce the damage to the hydrogel beads during the regeneration process, the agents used in the preparation process were selected to perform dye desorption experiments. The dye-adsorbed MHAF-SA and MHAF-CS hydrogel beads were immersed in 4.4 M acetic acid solution and 4.4 M ammonia solution, respectively. After shaking for 10 min, the hydrogel beads were washed several times and then dried at 50 °C for 6 h to obtain the regenerated hydrogel beads.

4. Conclusions

Husk of agarwood fruit-based hydrogel beads (MHAF-SA and MHAF-CS) were successfully synthesized in this study. The functional groups of the hydrogel beads were identified and proved by the FTIR spectra. The dynamic adsorption behavior of CV onto MHAF-SA and RB4 onto MHAF-CS can be represented well by the pseudo-second-order model and pseudo-first-order model, respectively. The adsorption isotherm data of CV onto MHAF-SA and RB4 onto MHAF-CS can be explained by the Langmuir isotherm model, with \( q_m \) values of 232.56–370.37 and 156.25–270.27 mg/g, respectively. Both hydrogel beads have excellent adsorption selectivity. After five adsorption-desorption cycles, the hydrogel beads still have a high removal efficiency. Overall, the hydrogel beads synthesized in this study have effectiveness, selectivity, and reusability and can be applied to the removal of dyes in aqueous solutions.

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References

1. Sahraei, R.; Pour, Z.S.; Ghaemy, M. Novel magnetic bio-sorbent hydrogel beads based on modified gum tragacanth/graphene oxide: Removal of heavy metals and dyes from water. J. Clean. Prod. 2016, 137, 189–194. [CrossRef]
2. Kurczewska, J.; Cegłowski, M.; Schroeder, G. Alginate/PAMAM dendrimer- Halloysite beads for removal of cationic and anionic dyes. Int. J. Biol. Macromol. 2019, 123, 398–408. [CrossRef]
3. Sun, H.; Cao, L.; Lu, L. Magnetite/reduced graphene oxide nanocomposites: One step solvolothermal synthesis and use as a novel platform for removal of dye pollutants. Nano Res. 2011, 4, 550–562. [CrossRef]
4. Garba, Z.N.; Zhou, W.; Lawan, I.; Xiao, W.; Zhang, M.; Wang, L.; Chen, L.; Yuan, Z. An overview of chlorophenols as contaminants and their removal from wastewater by adsorption: A review. J. Environ. Manag. 2019, 241, 59–75. [CrossRef] [PubMed]
18. Perez-Calderon, J.; Santos, M.V.; Zaritzky, N. Reactive RED 195 dye removal using chitosan coacervated particles as bio-sorbent:

16. Zhang, H.; Luan, Q.; Tang, H.; Huang, F.; Zheng, M.; Deng, Q.; Xiang, X.; Yang, C.; Shi, J.; Zheng, C.; et al. Removal of methyl blue dye from aqueous solution. 

17. Gan, L.; Li, H.; Chen, L.; Xu, L.; Liu, J.; Deng, A.; Mei, C.; Shang, S. Graphene oxide incorporated alginate hydrogel beads for the removal of various organic dyes and bisphenol A in water. 

19. Liu, S.; Li, L. Unique gelation of chitosan in an alkali/urea aqueous solution. 

20. Akter, S.; Islam, M.T.; Zuso, Z.M.; Khan, S.I. Agarwood production—a multidisciplinary field to be explored in Bangladesh. 

22. Shajahan, A.; Shankar, S.; Sathiyselvan, A.; Narayan, K.S.; Narayanan, V.; Kavyarasan, V.; Ignacimuthu, S. Comparative studies of chitosan and its nanoparticles for the adsorption efficiency of various dyes. 

23. Wang, W.; Zhao, Y.; Bai, H.; Zhang, T.; Ibarra-Galvan, V.; Song, S. Methylene blue removal from water using the hydrogel beads of poly(vinyl alcohol)-sodium alginate-chitosan-montmorillonite. 

24. Yu, F.; Cui, T.; Yang, C.; Dai, X.; Ma, J. k-Carrageenan/Sodium alginate double-network hydrogel with enhanced mechanical properties, anti-swelling, and adsorption capacity. 

25. Carneiro, P.A.; Fugivara, C.S.; Nogueira, R.F.P.; Boralle, N.; Zanoni, M.V.B. A comparative study on chemical and electrochemical degradation of Reactive Blue 4 dye. 

26. Mohanty, S.; Mouldick, S.; Maji, S.K. Adsorption/photodegradation of crystal violet (basic dye) from aqueous solution by hydrothermally synthesized titanate nanotubes (TNT). 

27. Carneiro, P.A.; Fugivara, C.S.; Nogueira, R.F.P.; Boralle, N.; Zanoni, M.V.B. A comparative study on chemical and electrochemical degradation of Reactive Blue 4 dye. 

28. Kong, Y.; Zhuang, Y.; Han, Z.; Yu, J.; Shi, B.; Han, K.; Hao, H. Dye removal by eco-friendly physically cross-linked double network polymer hydrogel beads and their functionalized composites. 

29. Zeng, Y.; Qiu, L.; Wang, K.; Yao, J.; Li, D.; Simon, G.P.; Wang, R.; Wang, H. Significantly enhanced water flux in forward osmosis desalination with polymer-graphene composite hydrogels as a draw agent. 

30. Oussalah, A.; Boukerroui, A.; Aichour, A.; Djellouli, B. Cationic and anionic dyes removal by low-cost hybrid alginate/natural bentonite composite beads: Adsorption and reusability studies. 

31. Chatterjee, S.; Lee, M.W.; Woo, S.H. Adsorption of congo red by chitosan hydrogel beads impregnated with carbon nanotubes.
33. Song, Y.; Tan, J.; Wang, G.; Zhou, L. Superior amine-rich gel adsorbent from peach gum polysaccharide for highly efficient removal of anionic dyes. *Carbohydr. Polym.* 2018, 199, 178–185. [CrossRef] [PubMed]
34. Wu, M.; Chen, W.; Mao, Q.; Bai, Y.; Ma, H. Facile synthesis of chitosan/gelatins filled with graphene bead adsorbent for orange II removal. *Chem. Eng. Res. Des.* 2019, 144, 35–46. [CrossRef]
35. Lagergren, S. Zur theorie der sogenannnten adsorption geloester stoffe, Kungliga Svenska Vetenskapsakad. *Handlingar* 1898, 24, 1–39.
36. Ho, Y.S.; McKay, G. Pseudo-second-order model for sorption process. *Process. Biochem.* 1999, 34, 451–465. [CrossRef]
37. Wang, J.; Liu, G.; Li, T.; Zhou, C. Physicochemical studies toward the removal of Zn(II) and Pb(II) ions through adsorption on montmorillonite-supported zero-valent iron nanoparticles. *RSC Adv.* 2015, 5, 29859–29871. [CrossRef]
38. Langmuir, I. The adsorption of gases on plane surfaces of glass, mica and platinum. *J. Am. Chem. Soc.* 1918, 40, 1361–1403. [CrossRef]
39. Freundlich, H.M.F. Über die absorption in lösungen. *Z. Phys. Chem.* 1906, 57, 385–470.
40. Chakraborty, S.; Chowdhury, S.; Das Saha, P. Adsorption of crystal violet from aqueous solution onto NaOH-modified rice husk. *Carbohydr. Polym.* 2011, 86, 1533–1541. [CrossRef]
41. Yan, H.; Li, H.; Yang, H.; Li, A.; Cheng, R. Removal of various cationic dyes from aqueous solutions using a kind of fully biodegradable magnetic composite microsphere. *Chem. Eng. J.* 2013, 223, 402–411. [CrossRef]
42. Zhang, L.; Zhang, H.; Guo, W.; Tian, Y. Removal of malachite green and crystal violet cationic dyes from aqueous solution using activated sintering process red mud. *Appl. Clay Sci.* 2014, 93–94, 85–93. [CrossRef]
43. Tahir, N.; Bhatti, H.N.; Iqbal, M.; Noreen, S. Biopolymers composites with peanut hull waste biomass and application for Crystal Violet adsorption. *Int. J. Biol. Macromol.* 2017, 94, 210–220. [CrossRef]
44. Zhou, Y.; Zhang, M.; Wang, X.; Huang, Q.; Min, Y.; Ma, T.; Niu, J. Removal of crystal violet by a novel cellulose-based adsorbent: Comparison with native cellulose. *J. Ind. Eng. Chem.* 2014, 53, 5498–5506. [CrossRef]
45. Bayramoglu, G.; Celik, G.; Arica, M.Y. Biosorption of reactive blue 4 dye by native and treated fungus phanerocheate chrysosporium: Batch and continuous flow system studies. *J. Hazard. Mater.* 2006, 137, 1689–1697. [CrossRef] [PubMed]
46. Binupriya, A.R.; Sathishkumar, M.; Jung, S.H.; Song, S.H.; Yun, S.I. A novel method in utilization of bokbunja seed wastes from wineries in liquid-phase sequestration of reactive blue 4. *Int. J. Environ. Res.* 2009, 3, 1–12.
47. Sun, D.; Zhang, Z.; Wang, M.; Wu, Y. Adsorption of reactive dyes on activated carbon developed from enteromorpha prolifera. *Am. J. Analyt. Chem.* 2013, 4, 17–26. [CrossRef]
48. Ma, C.M.; Hong, G.B.; Wang, Y.K. Performance Evaluation and Optimization of Dyes Removal using Rice Bran-Based Magnetic Composite Adsorbent. *Materials* 2020, 13, 2764. [CrossRef]
49. Zheng, X.; Li, X.; Li, J.; Wang, L.; Jin, W.; Liu, J.; Pei, Y.; Tang, K. Efficient removal of anionic dye (Congo red) by dialdehyde microfibrillated cellulose/chitosan composite film with significantly improved stability in dye solution. *Int. J. Biol. Macromol.* 2018, 107, 283–289. [CrossRef]
50. Weng, L.; Romanov, A.; Rooney, J.; Chen, W. Non-cytotoxic, in situ gelable hydrogels composed of N-carboxyethyl chitosan and oxidized dextran. *Biomaterials* 2008, 29, 3905–3913. [CrossRef] [PubMed]
51. Elwakeel, K.Z.; El-Kousy, S.; El-Shorbagy, H.G.; El-Ghaffar, M.A.A. Comparison between the removal of reactive black 5 from aqueous solutions by 3-amino-1,2,4 triazole,5-thiol and melamine grafted chitosan prepared through four different routes. *J. Environ. Chem. Eng.* 2016, 4, 733–745. [CrossRef]
52. Witono, J.R.; Noordergraaf, I.W.; Heeres, H.J.; Janssen, L.P. Water absorption, retention and the swelling characteristics of cassava starch grafted with polyacrylic acid. *Carbohydr. Polym.* 2014, 103, 325–332. [CrossRef] [PubMed]
53. Elwakeel, K.Z. Removal of reactive black 5 from aqueous solutions using magnetic chitosan resins. *J. Hazard. Mater.* 2009, 167, 383–392. [CrossRef] [PubMed]