Climate change risk to forests in China associated with warming

Yunhe Yin1, Danyang Ma1,2 & Shaohong Wu1,2

Variations in forest net primary productivity (NPP) reflects the combined effects of key climate variables on ecosystem structure and function, especially on the carbon cycle. We performed risk analysis indicated by the magnitude of future negative anomalies in NPP in comparison with the natural interannual variability to investigate the impact of future climatic projections on forests in China. Results from the multi-model ensemble showed that climate change risk of decreases in forest NPP would be more significant in higher emission scenario in China. Under relatively low emission scenarios, the total area of risk was predicted to decline, while for RCP8.5, it was predicted to first decrease and then increase after the middle of 21st century. The rapid temperature increases predicted under the RCP8.5 scenario would be probably unfavorable for forest vegetation growth in the long term. High-level risk area was likely to increase except RCP2.6. The percentage area at high risk was predicted to increase from 5.39% (2021–2050) to 27.62% (2071–2099) under RCP8.5. Climate change risk to forests was mostly concentrated in southern subtropical and tropical regions, generally significant under high emission scenario of RCP8.5, which was mainly attributed to the intensified dryness in south China.

Climate change has been suggested as a major driver of the dynamics of terrestrial ecosystems through its influence on vegetation growth and distribution1–4. Ecosystems may be at risk if their resilience and ability to adapt are severely damaged. An increase in global warming will result in greater risks for ecosystems over the 21st century5,6. For example, if the temperature increases by more than 3 °C, 44% of global terrestrial ecosystems risk conversion from carbon sinks to carbon sources7. Forest, the most complicated of all terrestrial ecosystems8, plays an important role in the carbon cycle and accounts for 49% of terrestrial gross primary production9. The annual gross carbon uptake by global forests equates to roughly half of the carbon emitted from fossil fuels10. Research into the spatial and temporal patterns of the climate change risk faced by forests is crucial to determine priority areas that should be targeted for action to manage the consequences of climate change.

Because of higher temperatures and incidental droughts, tree mortality and related dieback may increase, posing threats to carbon storage, biodiversity, and production in forests10–12. Net primary productivity (NPP) refers to the net amount of carbon fixed by plants through photosynthesis per unit time and area, namely the difference between gross primary productivity and autotrophic respiration13,14. Because the loss of NPP is considered to be unfavorable for terrestrial carbon sinks and ecosystem functioning, it has been used to indicate the risks of climate change for ecosystems6,15,16. In general, NPP is an indicator of plant growth and reflects the capacity of vegetation to sequester and convert the products of photosynthesis17–19. Changes in the NPP of terrestrial ecosystems could effectively reflect the substantial spatial and temporal heterogeneity in climatic, ecological, geochemical, and human influences on the biosphere17,20–22.

The impacts of climate change on forest NPP are complex. Warming and prolongation of the growing season may enhance forest NPP in high latitude and alpine areas23–25. In contrast, other factors such as drought, heat waves, wildfire, and insect disturbances may cause extensive reductions in NPP in some forests26–31. Of these factors, drought or dryness have often resulted from increases in the atmospheric evaporative demand (AED) and soil water deficit, and from decreases in precipitation. However, there is uncertainty about whether the current trends of regional aridity will intensify or weaken, which partly reflect AED estimation methods32–37. Generally, there is risk if the reduction in NPP caused by the overall effects of climate change indicates possible damage or corruption, such as poor vegetation coverage, the expansion of desertification38 and the decline of ecosystem services ability39, other than the lack of vegetation productivity and biomass. For example, based on version 2 of the Atmosphere Vegetation Interaction Model and the IPCC SRES B2 scenario, Shi et al.40 recently assessed the risks

1Key Laboratory of Land Surface Pattern and Simulation, Institute of Geographic Sciences and Natural Resources Research, Chinese Academy of Sciences, 11A, Datun Road, Chaoyang District, Beijing, 100101, China. 2University of Chinese Academy of Sciences, 19A Yuquan Road, Shijingshan District, Beijing, 100049, China. Correspondence and requests for materials should be addressed to Y.Y. (email: yinyh@igsnrr.ac.cn)
to ecosystems mainly including forest, shrubland, grassland and cultivated land in China from climate change when (1) the adverse variance was beyond the typical natural variability of NPP, and (2) there was the possibility that, in the future, the NPP would fall below the minimum NPP during the baseline period. However, we have little information regarding how the risks might respond to climate change, though it is important when assessing future impacts in forests in China. More importantly, there have been few quantitative assessments of the rates at which risk might change as regional mean temperatures increase.

The primary objectives of this study were therefore to (1) quantify the spatio-temporal patterns of different levels of risks indicated by the loss of forest NPP under future climate change in China, (2) investigate how the rate of regional climate change risks would vary in response to warming in the future, and (3) address the relative contributions of climatic factors to the risks. To achieve these objectives, we modeled NPP by modifying the AED sub-model in the Lund–Potsdam–Jena Dynamic Global Vegetation Model (LPJ-DGVM), which was driven by a set of general circulation models (GCMs) with representative concentration pathways (RCPs) that covered the period 1981–2100. Our research highlighted the different levels of risks to forests from climate change, and it should give us an improved understanding of how forest ecosystems will respond to future climate change.

Materials and Methods

Study area. Forests in China cover about 21.63% (2009–2013) of the land surface and have a carbon stock of approximately 84.27 billion tons. They account for a significant component of terrestrial ecosystems in China and represent the largest afforested area in the world41. Hou42 reported that needleleaf forest, the most widespread forest vegetation type, extended from the cold temperate zone to the tropical zone, and was mainly concentrated in southwestern and northeastern China. This forest type is dominated by Pinaceae, Taxodiaceae, and Cupressaceae. Mixed needleleaf and broadleaf forest is mainly found in northeastern China (e.g. Pinus koraiensis) and mountainous areas of southern China (e.g. Tsuga spp. and Chamaecyparis). Broadleaf forest is mainly concentrated in the eastern part of Northeast China, south of the Qinling Mountains, and in the southeastern part of the Tibetan Plateau.

We chose our study area from the distribution of forest types presented on the 1:1,000,000 vegetation map of China42 and the distribution of eco-geographical regions in China 43, as shown in Fig. 1. We simulated NPP from 1981 to 2099 at a spatial resolution of 0.5° × 0.5° with the modified LPJ-DGVM.

Risk assessment. Changes in ecosystem state are generally described by the anomaly of future projections from the historical long-term average, with the assumption that the standard deviation of the state variable represents the ecosystem natural interannual variability (IAV)5,16,44,45. NPP is defined as the rate of accumulation of carbon after losses from plant respiration and other metabolic processes that maintain the plant's living systems are taken into account46. In this study, we considered that there was risk when the absolute value of future negative anomaly of NPP exceeded the IAV for the period from 1981 to 2010, based on the LPJ simulations. The risks were ranked to correspond with the multiple relationships between the decreases in the NPP and the baseline IAV. Using the simulated NPP, we judged the risk at each pixel and counted the area of forest at risk based on smoothed data, i.e. 30-year running mean NPP. Linear trends in the time series of risk were detected by the Ordinary Least Squares (OLS) method; the statistical significance of the trends was determined by the non-parametric Mann-Kendall test.
Specifically, we calculated the anomalies of the forest NPP in China for future periods (from 2021 to 2050 and from 2071 to 2099) from the baseline period (from 1981 to 2010). Positive anomalies indicated no risk. Negative anomalies were compared with their baseline standard deviation to indicate different levels of risk for each pixel. In line with the criteria and indicators applied by Scholze et al. and Heyder et al., the risks were deemed low, medium, and high level when the absolute values of the negative NPP anomaly were less than half of the standard deviation ($\alpha < 0.5$), greater than half but less than one standard deviation ($0.5 < \alpha < 1$), or greater than one standard deviation ($\alpha > 1$), respectively. In this study, we paid more attention to abnormal variabilities in the NPP that were less than the average minus a number of standard deviations, as a certain loss of productivity outside the typical natural NPP variability was perceived to be an unacceptable impact from climate change. From this, we then classified the risk levels from the multiple relationships between the anomalies and standard deviation.

The standard deviation was computed as follows:

$$\delta = \sqrt{\frac{\sum_{i=1}^{n} (x_i - \bar{x})^2}{n-1}}$$

where $\delta$ was the standard deviation at the pixel scale, $x_i$ was the annual NPP of year $i$, $\bar{x}$ was the average NPP of the baseline period, and $n$ was the total number of baseline years ($n = 30$).

The multiple $\alpha$ was determined by the following equation:

$$\alpha = \frac{|x_i - \bar{x}|}{\delta}$$

The risk levels of forest productivity were classified from the above multiples.

**LPJ Model.** Forest NPP in China was projected with the process-based LPJ model. In the LPJ model, NPP is calculated by subtracting maintenance and growth respiration from GPP, which is computed by coupling the photosynthesis and water balance schemes. The formula is as follows:

$$\text{NPP} = 0.75(\text{GPP} - \text{R}_m)$$

where $\text{R}_m$ is the maintenance respiration and 0.75 is the ratio of NPP in the remainder, considering 0.25 is the growth respiration coefficient.

Proven as a useful tool for simulating the structure and function of large-scale ecosystems, the LPJ is driven by data of monthly climate and soil texture. The model comprises 10 plant functional types (PFTs), defined by bioclimatic limits and physiological optima, which compete for resources and determine vegetation composition. We used the method outlined by Zhao et al. to adjust a few of the PFT parameters, mainly including the maximum coldest monthly mean temperatures of boreal needle-leaved evergreen forest, boreal needle-leaved summergreen forest and boreal broad-leaved summergreen forest, to fit the characteristics of ecosystems in China.

The original method for calculating AED adds substantially to the uncertainty that is already associated with the climate change signal between GCMs. McVicar et al. advocated that all four primary meteorological variables, i.e., wind speed, atmospheric humidity, radiation, and air temperature, should be considered when assessing trends in AED. However, temperature-based methods, such as the Thornthwaite empirical equation, tend to underestimates AED. The Pennan–Monteith model recommended by the Food and Agricultural Organization (FAO56–PM model) reference crop evapotranspiration method gave good estimations of spatial and seasonal variability in AED across Great Britain. We therefore modified the original method for calculating AED in the evapotranspiration sub-model of the LPJ and used the physically-based FAO56–PM model, which had been previously calibrated for the study area.

The original fire module in the LPJ assumed only a minimum fuel load for fire spread. The model has been modified to include a maximum fuel load based on the linear relationship between the fuel load and the fire occurrence probability, so that the influence of fuel availability on fire occurrence is better reflected. Li et al. obtained good results when they applied a similar approach to the fire module of CLM-DGVM (Community Land Model with the Dynamic Global Vegetation Model). We expressed the probability of fire based on the available fuel as:

$$P = \max\left[0, \min\left\{1, \frac{L_{ag} - L_{low}}{L_{up} - L_{low}}\right\}\right]$$

where $L_{ag}$ (g C m$^{-2}$) is the fuel load (namely, the above-ground litter), and $L_{low}$ ($L_{low} = 200$) and $L_{up}$ ($L_{up} = 1000$) are the lower and upper fuel thresholds, respectively. Fire does not occur when $L_{ag} < L_{low}$ and becomes more likely to occur as the fuel increases when $L_{low} \leq L_{ag} < L_{up}$. The fire load no longer limits the fire spread when $L_{ag} > L_{up}$ when the other conditions are satisfied.

To ensure carbon pools and vegetation coverage were in a state of equilibrium, we ran the LPJ model for an initial period of 1000 years using the climate data for the reference period from 1981 to 2010 repeatedly. We then ran the simulation from 2011 to 2099 using climate change scenario data and the atmospheric CO$_2$ concentration for 2010.

**Climate Projections.** Projections of climate change under various scenarios are useful for predicting future changes to ecosystems and their response to global change. We used scenario analysis to assess risk development and the impacts of climate change. Because of the uncertainty associated with GCMs, we simulated forest NPP with multiple GCM projections and used the mean from multiple models to characterize the risk under different
emission scenarios. We used five GCMs in this study from the Coupled Model Intercomparison Project Phase 5, including HadGEM2-ES, IPSL-CM5A-LR, GFDL-ESM2M, MIROC-ESM-CHEM, and NorESM1-M. The GCM outputs were bias-corrected by the Inter-Sectoral Impact Model Intercomparison Project (ISI-MIP) and downscaled to a spatial resolution of 0.5°. The climate variables were the average, maximum and minimum temperatures, precipitation, surface downwelling shortwave radiation, near-surface wind speed, and relative humidity.

We used four representative concentration pathways, namely RCP2.6, RCP 4.5, RCP 6.0, and RCP8.5 scenarios, which indicated that radiative forcing levels of 2.6, 4.5, 6.0, and 8.5 Wm\(^{-2}\) would be reached by 2100, respectively. RCP8.5 was the highest emission scenario with a radiative forcing of around 8.5 Wm\(^{-2}\) in 2100, which was equivalent to an atmospheric CO\(_2\) concentration of about 1370 ppm. By the end of the 21st century, the temperatures in forest areas of China were projected to increase by between 1.68 °C under RCP2.6 and 6.29 °C under RCP8.5 relative to the average temperature in the baseline period of 1981–2010 (Fig. 2).

Results

**LPJ Model Evaluation.** The total NPP in China was estimated at 3.61 ± 0.13 Gt C a\(^{-1}\) during the baseline period of 1981–2010, which is consistent with the studies of Mao et al.\(^{65}\), Yuan et al.\(^{66}\), and Pan et al.\(^{67}\). The total NPP of forests in China for the past three decades was estimated at 1.40 ± 0.04 Gt C a\(^{-1}\), which is close to the results of Zhuang et al.\(^{68}\) and Ren et al.\(^{69}\). The 30-year averaged NPP varied spatially and gradually decreased from the southeast to the northwest (Fig. 3a); the vegetation type also varied and transitioned from woodland to grassland. Figure 3b showed the distribution of baseline standard deviation in NPP, which was basically opposite to the pattern of its multi-year mean value and was relatively higher in North China and Hainan Island. The simulated values essentially agreed with the field measurements of NPP for more than 700 sites spread across the different eco-geographical regions in China, published in the Global Primary Production Data Initiative Products database\(^{70}\) (Fig. 4). At each field site, a relative error could be estimated from the difference between simulated value and observed value, which was then divided by the observed value and expressed as a percentage. The average relative error over the whole country was 9.94%, which was acceptable and shows that the modified LPJ model gave satisfactory simulations of terrestrial NPP in China, and could be used to predict carbon cycling in ecosystems.

---

**Figure 2.** Temperature anomalies in forest areas of China during the 21st century relative to the baseline period (1981–2010). Solid lines indicate the five GCM ensemble means under the RCP scenarios, and the shading indicates one standard deviation of the ensemble mean. The time series were smoothed using a 10-year running mean.

**Figure 3.** Net primary productivity in China for 1981–2010 and its standard deviation distribution. (a) 30-year mean NPP. (b) NPP standard deviation. Note that black spots in (a) indicate field sites. The field sites data were obtained from the Global Primary Production Data Initiative Products database\(^{70}\). The figure was generated using ArcGIS 10.1 software (http://www.esri.com/).
Temporal change in risk in the future. Changes in the areas of climate change risk faced by forest systems in China under the four RCPs are shown in Fig. 5. There was a general decrease in the forest area at risk in China in the future, especially under the relatively low emission scenarios. Under RCP2.6, RCP4.5 and RCP6.0, the risk area was predicted to decrease noticeably in the first half of the 21st century and then fluctuate gently in the second half. For the highest emission scenario RCP8.5, the risk area was first predicted to decrease and then to increase from around the 2050s, which was mainly determined by the change in the high-level risk area. The total areas at risk between 2011 and 2099 were predicted to average 15.58%, 23.64%, 33.22% and 36.07% under RCP2.6, RCP4.5, RCP6.0, and RCP8.5, respectively. For the different risk levels, the forest area at medium risk was less than that the area at low risk and did not exceed 10% over most periods, while the area at high risk exhibited significant increasing trends except RCP2.6. Toward the end of the 21st century, the high-risk area was projected to increase...
Spatial change of risk in the future. To reflect the change in spatial patterns of future risk, two typical periods of 2021–2050 and 2071–2099 were selected in this study, representing mid and long terms. Results from the other periods could be found in Supplementary Figure S1-S4. The risk of climate change to the forest NPP in China is predicted to be aggravated for all scenarios from 2021 to 2050 (Fig. 6). The risk area is predicted to be mainly concentrated in the tropical humid and southern subtropical humid regions under the two relatively low emission scenarios (RCP2.6 and RCP4.5). For the higher emission scenarios (RCP6.0 and RCP8.5), this area is likely to extend northward to the mid-subtropical humid region and the northern subtropical humid region. The total risk areas for the four RCPs covered 21.14%, 31.82%, 42.61%, and 40.24% of the whole forest area (Table 1). Moreover, the low risk area was predicted to account for the largest part of the total risk area, while the high-risk area was predicted to account for the smallest portion. The forest area at low risk was highest under RCP6.0 and was distributed continuously through the areas south of the middle and lower reaches of the Yangtze River. Under RCP8.5, the risks to the forest in the southeast and southwest areas were mainly medium and high. The areas predicted to have medium and high risk under RCP6.0 were predicted to increase significantly from 11.76% and 1.19% to 14.35% and 5.39%, respectively, under RCP8.5.

The model predicted that the forest NPP risk pattern in China would be noticeably different for the period 2071–2099 (Fig. 7) than for 2021–2050, with the total risk areas accounting for 8.95%, 16.94%, 28.48%, and 46.17% under RCP2.6, RCP4.5, RCP6.0, and RCP8.5, respectively (Table 2). Decreases in the low and medium risk areas mainly explain the sharp decline in the total risk area predicted for the first three RCPs between the periods 2071–2099 and 2021–2050. Some of the low and medium risk areas were predicted to translate into high risk, though this mainly
affected areas in southern China. Furthermore, the high-risk area accounted for the largest percentage of the total risk area under RCP6.0 and RCP8.5, while the medium risk area accounted for the smallest portion. The area of forest at risk under RCP8.5 was predicted to be greater during 2071–2099 than 2021–2050 under RCP8.5, mainly because of a significant increase in the high-risk area. This high-risk area was predicted to extend from southern and central China to northeastern China, accounting for 27.62% of the total forest area.

Rate of change in risk in response to warming.

Figure 8a shows changes in the percentage of the area at risk in response to variations in the temperature anomalies predicted by the different GCMs under RCP8.5, processed by quadratic curve fitting. The corresponding rates of change, namely the first derivative of the functions shown in Fig. 8a, are shown in Fig. 8b. When the temperature increased by between $0.80 \pm 0.21 ^\circ C$ and $5.76 \pm 1.40 ^\circ C$ relative to the baseline ($9.76 \pm 0.08 ^\circ C$), the risk area was predicted to first decrease and then increase so that it covered half of the total forest area. The rate of change in the risk area percentage was predicted to first slow down and then accelerate, with the most obvious change predicted by the GFDL-ESM2M. The multi-model mean values show that the NPP risk area made up approximately 38% of the whole forest area in China when the temperature increased by about $3 ^\circ C$ under RCP8.5; for a temperature increase of about $6 ^\circ C$, the NPP risk area was predicted to extend to almost 55% of the forest area, with a rate of change of 13% $^\circ C^{-1}$.

Rate of change in risk in response to precipitation change.

We further analyzed the rate of change in risk in response to precipitation change. Results showed that the percentage of risk area was projected to decrease in the future, as the precipitation anomaly ranged from $-2.11 \pm 2.46 \%$ to $10.13 \pm 8.21 \%$ relative to the baseline period (Fig. 9a). For every additional percentage point increase in precipitation, the area at risk tended to decrease by 1.67% of the total forest area (Fig. 9b). In terms of the difference among GCMs, IPSL-CM5A-LR showed a

| risk level | RCP2.6 | RCP4.5 | RCP6.0 | RCP8.5 |
|------------|--------|--------|--------|--------|
| low risk   | 6.26   | 6.47   | 10.14  | 9.82   |
| medium risk| 2.37   | 6.04   | 7.55   | 8.74   |
| high risk  | 0.32   | 4.42   | 10.79  | 27.62  |
| total risk | 8.95   | 16.94  | 28.48  | 46.17  |

Table 2. Percentage of forest area at risk in China from 2071 to 2099 under different RCPs.
small range of precipitation change and a quick drop of NPP risk area; the area at risk was projected to turn from decreasing to increasing, when the precipitation increased by about 13% in the case of HadGEM2-ES.

**Driving factors.** We attributed the IAV of NPP to different climatic factors by calculating the partial correlation coefficients between NPP and precipitation (PRE), temperature (TEM), and aridity index (AED/PRE) for the period 2071–2099 under RCP8.5 (Fig. 10). There were significant positive correlations between temperature and forest NPP, mainly in northeastern China and in the eastern part of the Tibetan Plateau, which occupied 9.60% of the total forest area. Significant negative correlations between precipitation and forest NPP accounted for a larger proportion (15.53%) and were mainly limited to southern China. Substantial temperature rise was projected to occur in Northeast China, whereas the relatively slight increase of precipitation seemed to be distributed in Southeast China. The aridity index tended to increase obviously for these areas, implying potential droughts in the long term under high emission scenario. The aridity index and NPP were significantly and negatively correlated for nearly half of the forests (49.84%), ranging from northeastern China to southern China, indicating that increased dryness would limit forest growth in these regions. Therefore, of the three variables, the projected aridity index seemed to have a greater impact on the forest NPP, both in its degree and extent, which means that risks of a decrease in the forest NPP were mainly because of dryness.

**Discussion**

The pattern predicted in this study is spatially consistent with other recent findings. For example, Gang et al.\(^7^1\) suggested that the terrestrial NPP would increase in the northern mid and high latitudes where warming would favor tree growth and expansion by the end of the 21st century, especially for temperate and boreal forests. Using the RCP4.5 scenario in the Integrated Biosphere Simulator, Yuan et al.\(^7^3\) found that the NPP of deciduous broad-leaved forest in the warm temperate zone in China would increase, while that of subtropical evergreen broad-leaved forests in southern China would decrease, and would be vulnerable from 2016 to 2050.
We found that the total area of forest in China at risk from climate change would first decrease and then increase as the warming accelerated under RCP8.5. The future projections of forest risk to climate change indicate that climate warming may be beneficial to vegetation growth to a minor extent, but when the climate change exceeds a certain threshold, the impacts could be negative. While increases in temperature could enhance plant photosynthesis, they could also cause a water vapor pressure deficit. The leaf stomata would then close to prevent water loss and increase water use efficiency. Several previous studies have reported that, without considering the CO₂ fertilization effect, rapid temperature increases and increased frequency of drought events could cause a decrease in the NPP in tropical and subtropical forests, and even in the global terrestrial NPP, under relatively high emission scenarios. Because of warming and changes in rainfall patterns, the decrease in the available soil water has slowed down the increase in the forest NPP in southern China over the past three decades. From their analysis of 32 years of data from forest observation plots, Zhou et al. suggested that subtropical forests in China were threatened by their lack of resilience to long-term climate change manifested by rising temperatures and increased occurrence of soil drying. Also, studies of the impact of climate change on aridity during the 21st century have predicted increased aridity over most tropical and mid-latitude land regions, and, in particular, over

Figure 10. Anomalies of climatic factors during 2071–2099 relative to 1981–2010 under RCP8.5 and their partial correlation coefficients with forest NPP in China. (a,b) temperature. (c,d) precipitation. (e,f) aridity index. Note that black checkmarks indicate statistical significance (p < 0.05). The figure was generated using ArcGIS 10.1 software.
most of Africa, the Americas, Australia, Southeast Asia, and the Mediterranean region. This indicates that more intense droughts would limit future forest growth in low latitude regions. Climate change alone may lead to less overall tree coverage in the Tropics, while the competing effects of CO2 fertilization and climate change, along with the uncertainty of projected precipitation changes in the Tropics, mean that there is a large degree of uncertainty associated with projected future changes in vegetation9.

Moreover, shifts in forest disturbances such as wind, pests, and fire may adversely affect forest productivity under future climate change. The productivity of and carbon storage in Europe’s forests is likely to decrease as climate change and forest disturbances intensify21,50, though both increasing and decreasing trends have been found in the growth and productivity across Europe. For example, the growth of European beech was observed to have declined because of droughts over the past 20th century46, and its NPP and water-use efficiency was predicted to reduce under future climate conditions (A1B scenario) due to aggravated water shortage and droughts.44. As the risks from forest fire are predicted to become increasingly serious in China, mainly because of an increase in fire weather in central and southeastern China48, measures should be implemented to reduce the negative impacts of fire disturbance on forest productivity. The sensitivity of NPP to climate change is also the key to understanding how risk develops and evolves. Piao et al.21 found that the inter-annual correlation between terrestrial productivity and temperature decreased in temperate regions because of an increase in drought over the past decade, while Heyder et al.46 reported that strong warming could amplify the sensitivity to declining precipitation in temperate and tropical ecosystems.

There are uncertainties in risk forecasts in vegetation productivity from several sources, including emission scenarios, climate models, and ecological models. For instance, due to the large discrepancy in future projections of precipitation from global to regional scale46,47, the difference in precipitation patterns and extreme events across CMIP5 models could be a vital source of uncertainty for terrestrial carbon flux and its impact from climate change.88. As for China, the cross-model variability of future NPP was reported to be significantly contributed by the simulated precipitation on the local scale, especially in northwestern area.88. From their assessment, Sitch et al.92 found that the responses of five DGVMs to climate change varied more widely than their responses to changes in CO2 concentrations. Nishina et al.93 considered that, if the uncertainty in the ISI-MIP results were to be reduced, the simulation capacity of vegetation models would need to improve. When estimating forest biomass and productivity, accurate descriptions and determinations of allometry and allometric scaling parameters, respectively, are important85. We did not consider the direct effect of CO2 fertilization in our study even though it is an important influence on changes in vegetation NPP. The observed increase in photosynthetic water-use efficiency in temperate and boreal forests of the Northern Hemisphere over the past two decades has been closely associated with elevated atmospheric CO2 concentrations.85. Water stress or decreased carbon gains from autotrophic respiration may result in decreased vegetation productivity and loss of forest cover when there are short-ages of CO2.84. There was a considerable difference between the simulated and observed data in the forest NPP in the tropical humid region for the baseline period, which probably reflects the fact that there were only seven site records from this area, thereby causing high uncertainty. To measure this kind of uncertainty, the time series of observed NPP from 1981 to 2010 are needed to compute the standard deviation and to further calculate the risk. Nevertheless, due to the lack of time series data in the observed NPP across China, we are unable to quantify the uncertainty resulting from the difference between simulations and observations. Further, different ways of assessing risk may also create uncertainties. For example, baseline values may differ depending on whether they are derived from the global average or regional average, and variable values may exceed natural variability when they are beyond the mean plus or minus several standard deviations.5,16,47. All these factors can produce different risk assessment results. van Oijen98,99 defined risk as the product of probability and vulnerability, and used the difference in NPP between hazardous condition and normal climate to indicate ecosystem vulnerability. Although the method is explicit in mathematics and easy to operate, it seems more suitable to the risk induced by single hazard factor and the distinction between hazardous and non-hazardous conditions mainly depended on the subjective experience. In addition to the impact of climate change, human activities like forest management measures may also drive changes in ecosystem100. It deserves more research on how to isolate the climate change risk from its complicated interactions with human factors.

**Conclusion**

In this study, we modified the AED and fire modules of the LPJ model and investigated spatial and temporal features of climate change risk faced by forest in China in the 21st century under different RCPs. The risk rating predicted in this study from the relationships between decreases in NPP and baseline variability (indicated by the standard deviation) highlighted the adverse impact of future climate change on vegetation productivity. The area of forest NPP at risk in China showed a general tendency to decrease from 2011 to 2099 relative to the baseline period of 1981–2010, under RCP2.6, RCP4.5, and RCP6.0. High-level risk area would increase especially in RCP4.5 (0.7% per decade, p < 0.01) and RCP6.0 (1.7% per decade, p < 0.01).

The risk of climate change to forest in China is likely to be relatively obvious under RCP8.5 compared with low emission scenarios especially in the long term. In response to future climate change, the total risk area is predicted to first decrease and then increase after the middle of 21st century. The percentage area at high risk was predicted to increase from 5.39% (2021–2050) to 27.62% (2071–2099) with a trend of 3.2% per decade. The forest vegetation growth would probably be weakened as the degree of warming increased under RCP8.5.

Spatial distributions show that climate change risk to forests was projected to be concentrated in south China. The risk of future climate change to forest in China is predicted to be mainly distributed in the low latitude southern subtropical humid and tropical humid regions where there were intensified dryness and where the declines in productivity superimposed by natural hazards such as droughts and floods may bring huge losses to the local economy.
References

1. Nemani, R. R. et al. Climate-driven increases in global terrestrial net primary production from 1982 to 1999. Science 300, 1560–1563 (2003).

2. Seddon, A. W., Macias-Fauria, M., Long, P. R., Benz, D. & Willis, K. J. Sensitivity of global terrestrial ecosystems to climate variability. Nature 531, 229–232 (2016).

3. Zhao, M. S. & Running, S. W. Drought-induced reduction in global terrestrial net primary production from 2000 through 2009. Science 329, 940–943 (2010).

4. Pearson, R. G. et al. Shifts in Arctic vegetation and associated feedbacks under climate change. Nat. Clim. Chang. 3, 673–677 (2013).

5. Scholze, M., Knorr, W., Arnell, N. W. & Prentice, I. C. A climate-change risk analysis for world ecosystems. Proc. Natl. Acad. Sci. USA 103, 13116–13120 (2006).

6. Warzacki, L. et al. A multi-model analysis of risk of ecosystem shifts under climate change. Environ. Res. Lett. 8, 044018 (2013).

7. FAO. Global Forest Resources Assessment2015. FAO Forestry Paper No. 1. (United Nations Food and Agriculture Organization, Rome, 2015).

8. Yan, Y. et al. A large and persistent carbon sink in the world’s forests. Science 333, 988–993 (2011).

9. IPCC. Summary for policymakers. Climate Change 2014: Impacts, Adaptation, and Vulnerability. Part A: Global and Sectoral Aspects. Contribution of Working Group II to the Fifth Assessment Report of the Intergovernmental Panel on Climate Change. 1–32 (Cambridge University Press, Cambridge, United Kingdom and New York, NY, USA, 2014).

10. Ramming, A. et al. Estimating the risk of Amazonian forest dieback, New Phytol. 187, 694–706 (2010).

11. Ciais, P. et al. Europe-wide reduction in primary productivity caused by the heat and drought in 2003. Nature 437, 529–533 (2005).

12. Field, C. B., Behrenfeld, M. J., Randerson, J. T. & Falkowski, P. Primary production of the biosphere: integrating terrestrial and oceanic components. Science 281, 237–240 (1998).

13. Melillo, J. M. et al. Global climate change and terrestrial net primary production. Nature 363, 234–240 (1993).

14. Yin, Y. Y., Tang, Q. H., Wang, L. X. & Liu, X. C. Risk and contributing factors of ecosystem shifts under naturally vegetated land under climate change in China. Sci. Rep. 6, 20905 (2016).

15. Heyder, U., Schaphoff, S. & Lucht, W. Risk of severe climate change impact on the terrestrial biosphere. Environ. Res. Lett. 6, 034036 (2011).

16. Cramer, W. et al. Global response of terrestrial ecosystem structure and function to CO2 and climate change: results from six dynamic global vegetation models. Glob. Chang. Biol. 7, 357–373 (2001).

17. Steffen, W. et al. The terrestrial carbon cycle: Implications for the Kyoto Protocol. Science 280, 1393–1394 (1998).

18. Wu, Z. et al. Shifts in Arctic vegetation and associated feedbacks under climate change: a meta-analysis of experimental manipulation. Glob. Chang. Biol. 17, 927–942 (2011).

19. Boivin, C. & Running, S. W. Impacts of climate change on natural forest productivity - evidence since the middle of the 20th century. Glob. Chang. Biol. 12, 862–882 (2006).

20. Piao, S. et al. Evidence for a weakening relationship between interannual temperature variability and northern vegetation activity. Nat. Commun. 5, 5018 (2014).

21. Räfique, R., Zhao, F., Jong, R. D., Zeng, N. & Asrar, G. R. Global and regional variability and change in terrestrial ecosystems net primary production and NDVI: A model-data comparison. Remote Sens. 8, 177 (2016).

22. Andreu-Hayles, L. et al. Varying boreal forest response to Arctic environmental change at the Firth River, Alaska. Environ. Res. Lett. 6, 045503 (2011).

23. Euskirchen, E. S., McGuire, A. D., Chapin III, F. S., Yi, S. & Thompson, C. C. Changes in vegetation in northern Alaska under scenarios of climate change, 2003–2100: implications for climate feedbacks. Ecol. Appl. 19, 1022–1043 (2009).

24. Piao, S. et al. Changes in vegetation net primary productivity from 1982 to 1999 in China. Global Biogeochem. Cycles 19, 1605–1622 (2005).

25. Allen, C. D. et al. A global overview of drought and heat-induced tree mortality reveals emerging climate change risks for forests. Forest Ecol. Manage. 259, 660–684 (2010).

26. Littell, J. S. et al. Forest ecosystems, disturbance, and climatic change in Washington State, USA. Clim. Chang. 102, 129–158 (2010).

27. Murray-Tortolero, G. et al. The dry season intensity as a key driver of NPP trends. Geophys. Res. Lett. 43, 2632–2639 (2016).

28. Wang, J. et al. Decreasing net primary production due to drought and slight decreases in solar radiation in China from 2000 to 2012. J. Geophys. Res. Biogeosci 112, 261–278 (2017).

29. Fung, I. Y., Doney, S. C., Lindsay, K. & John, J. Evolution of carbon sinks in a changing climate. Proc. Natl. Acad. Sci. USA 102, 11201–11206 (2005).

30. Seidl, R., Schelhaas, M. J., Rammer, W. & Verkerk, P. J. Increasing forest disturbances in Europe and their impact on carbon storage. Nat. Clim. Chang. 4, 806–810 (2014).

31. Ghee, P. et al. Global assessment of trends in wetting and drying over land. Nature Geosci. 7, 716–721 (2014).

32. Mueller, B. & Zhang, X. B. Causes of drying trends in northern hemispheric land areas in reconstructed soil moisture data. Clim. Chang. 134, 255–267 (2016).

33. Trenberth, K. E. et al. Global warming and changes in drought. Nat. Clim. Chang. 4, 17–22 (2014).

34. Dai, A. Global increasing drought under global warming in observations and models. Nat. Clim. Chang. 3, 52–58 (2013).

35. Sheffield, J., Wood, E. F. & Doderick, M. L. Little change in global drought over the past 60 years. Nature 491, 455–458 (2012).

36. Feng, S. & Fu, Q. Expansion of global drylands under a warming climate. Atmos. Chem. Phys. 13, 10081–10094 (2013).

37. Xu, D. Y., Kang, X. W., Zhang, D. F. & Pan, J. J. Multi-scale quantitative assessment of the relative roles of climate change and human activities in desertification - A case study of the Ordos Plateau, China. Journal of Arid Environments 74, 498–507 (2010).

38. Fu, Y., Lu, X., Zhao, Y., Zeng, X. & Xia, J. Assessment impacts of weather and land use/land cover (LULC) change on urban vegetation net primary productivity (NPP): A case study in Guangzhou, China. Remote Sens. 5, 4121–4144 (2013).

39. Shi, X. et al. Climate change risks for net primary production of ecosystems in China. Hum. Ecol. Risk Assess. 22, 1091–1105 (2016).

40. State Forestry Administration. The eighth national forest resource inventory results. Forest Resources Management (in Chinese), 1–2 (2014).

41. Hou, X. Vegetation Atlas of China. (Science Press, 2001).

42. Zheng, D. A Study on the Eco-Geographic Regional System of China. (FAO FRA2000 Global Ecological Zoning Workshop, Cambridge, UK, 1999).

43. Xu, M. et al. Effects of climatic factors and ecosystem responses on the inter-annual variability of evapotranspiration in a confiuseous plantation in subtropical China. Plos One 9, e85593 (2014).

44. Yuan, W. et al. Latitudinal patterns of magnitude and interannual variability in net ecosystem exchange regulated by biological and environmental variables. Glob. Chang. Biol. 15, 2905–2920 (2009).

45. Lieth, H. & Whittaker, R. H. Primary Productivity of the Biosphere. (Springer-Verlag, 1975).

46. van Minnen, J. G., Onigkeit, J. & Alcamo, I. Critical climate change as an approach to assess climate change impacts in Europe: development and application. Environ. Sci. Technol. 5, 335–347 (2002).
48. Sitch, S. et al. Evaluation of ecosystem dynamics, plant geography and terrestrial carbon cycling in the LPJ dynamic global vegetation model. Glob. Chang. Biol. 9, 161–185 (2003).

49. Zhao, D., Wu, S. & Yin, Y. Responses of terrestrial ecosystems’ net primary productivity to future regional climate change in China. Plos One 8, e60649 (2013).

50. Kingston, D. G., Todd, M. C., Taylor, R. G., Thompson, J. R. & Arnell, N. W. Uncertainty in the estimation of potential evapotranspiration under climate change. Geophys. Res. Lett. 36, L20403 (2009).

51. McVicar, T. R. et al. Global review and synthesis of trends in observed terrestrial near-surface wind speeds: Implications for evaporation. J. Hydrol. 416-417, 182–205 (2012).

52. Trajkovic, S. Temperature-based approaches for estimating reference evapotranspiration. J. Irrig. Drain. E-Asce 131, 316–323 (2005).

53. Donohue, R. J., McVicar, T. R. & Roderick, M. L. Assessing the ability of potential evaporation formulations to capture the dynamics in evaporative demand within a changing climate. J. Hydrol. 386, 186–197 (2010).

54. Prudhomme, C. & Williamson, J. Derivation of RCM-driven potential evapotranspiration for hydrological climate change impact analysis in Great Britain: a comparison of methods and associated uncertainty in future projections. Hydrol. Earth Syst. Sci. 17, 1365–1377 (2013).

55. Jarvis, P. G. & McNaughton, K. G. Stomatal Control of Transpiration: Scaling Up from Leaf to Region. Adv. Ecol. Res. 15, 1–49 (1986).

56. Allen, R. G., Pereira, L. S., Raes, D. & Smith, M. Crop Evapotranspiration: Guidelines for Computing Crop Water Requirements, FAO Irrigation and Drainage paper 56. (United Nations Food and Agriculture Organization, 1998).

57. Yin, Y. H., Wu, S. H., Zheng, D. & Yang, Q. Y. Radiation calibration of FAO56 Penman-Monteith model to estimate reference crop evapotranspiration in China. Agric. Water Manage. 95, 77–84 (2008).

58. Thonickie, K., Venevsky, S., Sitch, S. & Cramer, W. The role of fire disturbance for global vegetation dynamics: coupling fire into a Dynamic Global Vegetation Model. Glob. Ecol. Biogeogr. 10, 661–677 (2001).

59. Arora, V. K. & Boer, G. J. Fire as an interactive component of dynamic vegetation models. J. Geophys. Res. 110, 149–167 (2005).

60. Li, F., Zeng, X. & Levis, S. A process-based fire parameterization of intermediate complexity in a Dynamic Global Vegetation Model. Biogeosciences 9, 2761–2780 (2012).

61. Taylor, K. E., Stouffer, R. J. & Meehl, G. A. An Overview of CMIP5 and the Experiment Design. Bull. Amer. Meteor. Soc. 93, 485–498 (2012).

62. Hempel, S., Frieler, K., Warszawski, L., Schewe, J. & Piontek, F. A trend-preserving bias correction - the ISI-MIP approach. Earth Syst. Dynam. 4, 219–236 (2013).

63. Warszawski, L. et al. The Inter-Sectoral Impact Model Intercomparison Project (ISI-MIP): Project framework. Proc. Natl. Acad. Sci. USA 111, 3228–3232 (2013).

64. Moss, R. H. et al. The next generation of scenarios for climate change research and assessment. Nature 463, 747–756 (2010).

65. Mao, J., Dan, L., Wang, B. & Dai, Y. Simulation of terrestrial ecosystem NPP with M-SDGVM over Continental China. Adv. Atmos. Sci. 27, 427–442 (2010).

66. Yuan, Q. et al. Modeling net primary productivity of the terrestrial ecosystem in China from 1961 to 2005. J. Geogr. Sci. 24, 3–17 (2014).

67. Pan, S., Tian, H., Lu, C., Dangal, S. R. S. & Liu, M. Net primary production of major plant functional types in China: Vegetation classification and ecosystem simulation. Acta Ecol. Sinica 35, 28–36 (2015).

68. Zhuang, Q., Zhang, T., Xiao, J. & Luo, T. Quantification of net primary production of Chinese forest ecosystems with spatial statistical approaches. Mitig. Adapt. Strateg. Glob. Chang. 14, 85–99 (2009).

69. Ren, W. et al. Impacts of tropospheric ozone and climate change on net primary productivity and net carbon exchange of China’s forest ecosystems. Glob. Ecol. Biogeogr. 20, 391–406 (2011).

70. Olson, R. J., Sculrock, J. M. O., Prince, S. D., Zheng, D. L. & Johnson, K. R. NPP Multi-Biome: Global Primary Production Data Initiative Products, R2. (ORNL Distributed Active Archive Center, Oak Ridge, Tennessee, USA, 2013).

71. Gang, C. et al. Modeling the dynamics of distribution, extent, and NPP of global terrestrial ecosystems in response to future climate change. Glob. Planet. Chang. 148, 153–165 (2017).

72. Yuan, Q. et al. NPP vulnerability of the potential vegetation of China to climate change in the past and future. J. Geogr. Sci. 27, 131–142 (2017).

73. Malhi, Y. et al. Exploring the likelihood and mechanism of a climate-change-induced dieback of the Amazon rainforest. Proc. Natl. Acad. Sci. USA 106, 20610–20615 (2009).

74. Pan, S. et al. Complex spatiotemporal responses of global terrestrial primary production to climate change and increasing atmospheric CO2 in the 21st century. Plos One 9, e112810 (2014).

75. Wang, W., Wang, J., Liu, X., Zhou, G. & Yan, J. Decadal drought deaccelerated the increasing trend of annual net primary production in tropical or subtropical forests in southern China. Sci. Rep. 6, 28640 (2016).

76. Zhou, G. et al. A climate change-induced threat to the ecological resilience of a subtropical monsoon evergreen broad-leaved forest in Southern China. Glob. Chang. Biol. 19, 1197–1210 (2013).

77. Gao, X. & Giorio, F. Increased aridity in the Mediterranean region under greenhouse gas forcing estimated from high resolution simulations with a regional climate model. Glob. Planet. Chang. 62, 195–209 (2008).

78. Dai, A. Drought under global warming: a review. WIREs Clim Change 2, 45–65 (2011).

79. Yu, M., Wang, G., Parr, D. & Ahmed, K. F. Future changes of the terrestrial ecosystem based on a dynamic vegetation model driven by RCP4.5 climate projections from 19 GCMS. Clim. Chang. 127, 257–271 (2014).

80. Rötzer, L., Görgen, Schuler & Pretzsch. Modelling the impact of climate change on the productivity and water-use efficiency of a central European beech forest. Clim. Res. 58, 81–95 (2013).

81. Ruiz-Benito, P. et al. Stand structure and recent climate change constrain stand basal area change in European forests: a comparison across boreal, temperate, and Mediterranean biomes. Ecosystems 17, 1439–1454 (2014).

82. Lindner, M. et al. Climate change and European forests: What do we know, what are the uncertainties, and what are the implications for forest management? Journal of Environmental Management 146, 69–83 (2014).

83. Ding, H., Chibabi, A., Silvestri, S. & Nunes, P. A. L. D. Valuing climate change impacts on European forest ecosystems. Ecosystem Services 18, 141–153 (2016).

84. Kint, V. et al. Radial growth change of temperate tree species in response to altered regional climate and air quality in the period 1901–2008. Clim. Chang. 115, 343–363 (2012).

85. Tian, X., Zhao, F., Shu, L. & Wang, M. Changes in forest fire danger for south-western China in the 21st century. Int. J. Wildland Fire 23, 185–195 (2014).

86. Hosseinzadehatalaei, P., Tabari, H. & Willems, P. Uncertainty assessment for climate change impact on extreme precipitation: how many model runs do we need? Int. J. Climatol. 37, 1105–1117 (2017).

87. Zhao, L., Xu, J., Powell Jr, A. M. & Jiang, Z. Uncertainties of the global-to-regional temperature and precipitation simulations in CMIP5 models for past and future 100 years. Theor. Appl. Climatol. 122, 259–270 (2015).

88. Peng, J. & Dan, L. Impacts of CO2 concentration and climate change on the terrestrial carbon flux using six global climate–carbon coupled models. Ecol. Model. 304, 69–83 (2015).
89. Wang, T., Lin, X., Liu, Y., Dantec-Nédélec, S. & Ottlé, C. Causes of uncertainty in China's net primary production over the 21st century projected by the CMIP5 Earth system models. *Int. J. Climatol.* **36**, 2323–2334 (2016).

90. Sitch, S. *et al.* Evaluation of the terrestrial carbon cycle, future plant geography and climate-carbon cycle feedbacks using five Dynamic Global Vegetation Models (DGVMs). *Glob. Chang. Biol.* **14**, 2015–2039 (2008).

91. Nishina, K. *et al.* Decomposing uncertainties in the future terrestrial carbon budget associated with emission scenario, climate projection, and ecosystem simulation using the ISI-MIP result. *Earth Syst. Dynam.* **6**, 435–445 (2015).

92. Wolf, A. *et al.* Forest biomass allometry in global land surface models. *Global Biogeochem. Cycles* **25**, GB3015 (2011).

93. Keenan, T. F. *et al.* Increase in forest water-use efficiency as atmospheric carbon dioxide concentrations rise. *Nature* **499**, 324–327 (2013).

94. Poulter, B. *et al.* Recent trends in Inner Asian forest dynamics to temperature and precipitation indicate high sensitivity to climate change. *Agr. Forest Meteorol.* **178-179**, 31–45 (2013).

95. Ni, J. Net primary productivity in forests of China: scaling-up of national inventory data and comparison with model predictions. *Forest Ecol. Manage.* **176**, 485–495 (2003).

96. Ni, J., Zhang, X. S. & Scurlock, J. M. O. Synthesis and analysis of biomass and net primary productivity in Chinese forests. *Ann. For. Sci.* **58**, 351–384 (2001).

97. Zhao, M. & Zhou, G. S. Estimation of biomass and net primary productivity in major planted forests in China based on forest inventory data. *Forest Ecol. Manage.* **207**, 295–313 (2005).

98. van Oijen, M. *et al.* A novel probabilistic risk analysis to determine the vulnerability of ecosystems to extreme climatic events. *Environ. Res. Lett.* **8**, 1880–1885 (2013).

99. van Oijen, M. *et al.* Impact of droughts on the C-cycle in European vegetation: a probabilistic risk analysis using six vegetation models. *Biogeosciences* **11**, 6357–6375 (2014).

100. Kröel-Dulay, G. *et al.* Increased sensitivity to climate change in disturbed ecosystems. *Nat. Commun.* **6**, 6682 (2015).

Acknowledgements

This study was supported by the National Natural Science Foundation of China (No. 41571043, 41530749), the National Scientific Technical Supporting Program of China (No. 2012BAC19B02), and the Cultivate Project of Institute of Geographic Sciences and Natural Resources Research, Chinese Academy of Sciences (No. TSYJS03). The authors thank the ISI-MIP for providing us with bias corrected climate data. We acknowledge the World Climate Research Programme’s Working Group on Coupled Modeling, which is responsible for CMIP, and we thank the climate modelling groups (GFDL-ESM2M, HadGEM2-ES, IPSL-CM5A-LR, MIROC-ESM-CHEM, and NorESM1-M) for producing and making available their model output.

Author Contributions

Y.Y. designed the study. D.M. conducted the analysis. Y.Y. and D.M. contributed to the paper writing. S.W. contributed suggestions for this paper.

Additional Information

Supplementary information accompanies this paper at https://doi.org/10.1038/s41598-017-18798-6.

Competing Interests: The authors declare that they have no competing interests.

Publisher’s note: Springer Nature remains neutral with regard to jurisdictional claims in published maps and institutional affiliations.

Open Access This article is licensed under a Creative Commons Attribution 4.0 International License, which permits use, sharing, adaptation, distribution and reproduction in any medium or format, as long as you give appropriate credit to the original author(s) and the source, provide a link to the Creative Commons license, and indicate if changes were made. The images or other third party material in this article are included in the article’s Creative Commons license, unless indicated otherwise in a credit line to the material. If material is not included in the article’s Creative Commons license and your intended use is not permitted by statutory regulation or exceeds the permitted use, you will need to obtain permission directly from the copyright holder. To view a copy of this license, visit http://creativecommons.org/licenses/by/4.0/.

© The Author(s) 2017