Effect of Different CO₂ Treatments on the Metal Leaching in Steel Slag Binders

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Carbonation is an effective method to promote the quality of the steel slag binder. In this article, two carbonation approaches, namely hot-stage carbonation and accelerated carbonation, were employed to leach the metals, and the influence mechanism on the metal sequential leachability of the binders composed of 80 wt% of EAF slag incorporating 20 wt% of Portland cement (PC) was revealed. The carbonate products, microstructures, and chemical states were investigated, and the results indicated that chromium, vanadium, and titanium gradually transformed into inactive phases after two carbonation approaches, while zinc appeared the opposite trend. The sequential leachability of chromium declined with the increase of the carbonation efficiency, in which the exchangeable chromium decreased from 1.99 mg/kg in the A2A binder to below the detection limit in the A2C binder and C2C binder. Hot-stage carbonation treatment facilitated particle agglomeration, minerals remodeling, and calcite formation. The carbonation curing of the steel slag paste resulted in the formation of amorphous CaCO₃, calcite crystalline and Si-bearing hydrates that covered the pores of the matrix, and silicate structure with a higher disorder. The hot-stage carbonation and accelerated carbonation curing methods were adopted to jointly prevent the leaching of harmful metals and facilitate promising high-volume steel slag-based binders with structural densification and CO₂ storage.

Keywords: accelerated carbonation curing, CO₂ storage, sequential leachability, steel slag, hot-stage carbonation

INTRODUCTION

Steel slags, the inevitable industrial waste in the process of steel manufacturing, account for 15–35% of crude steel (Eloneva et al., 2010). In China, the annual output of steel slags exceeds 100 million tons (Guo et al., 2018), but the utilization ratio is below 30% (Pan et al., 2017). Steel slags are mainly recycled for the fields of cement and concrete (Carvalho et al., 2017), road building (Pasetto and Baldo, 2016), bituminous mixture (Skaf et al., 2017), soil amelioration (Poh et al., 2006), and phosphate fertilizer (Yi et al., 2012). However, the utilization of the steel slag was restricted by the slow hydration and the undesired deleterious expansion due to the mineralogical changes, such as the hydration of the free CaO and MgO and the transformation of α-C₂S (Wang et al., 2010; Mo et al., 2017).

The carbonation of steel slags was proved as an effective method to solve these issues, which was due to the slower hydration reaction of steel slag than that of the carbonization in the humid environment and the improvement of mineralogical stability by the carbonization of the alkaline oxides (Ukwattage et al., 2017). Meanwhile, the carbonation of steel slags contributed to the storage of CO₂. Several researchers investigated the potential application of steel slags for carbon
sequestration to produce qualified building materials (Polettini and Pomi, 2004; Baciocchi et al., 2015; Kim et al., 2016; Pan et al., 2017). Moreover, hot-stage CO₂ sequestration not only provided recovery potential but also reused the waste heat from the exhaust gas to accelerate reactions, which could be a finishing step in the industrial plants (Pan et al., 2016).

According to previous studies, carbonation changed the performance of the steel slag, and much more attention should be paid to the leaching of hazardous elements (e.g., Cr, V, and Mo) (Belhadj et al., 2012). The steel slags from the stainless-steel production contain a lower FeO content but a higher Cr content (Shi, 2004). There were only limited studies about the mechanism between the carbonation and the leaching of hazardous elements. Kim et al. (Kim et al., 2016) stated that the leachability of Cr decreased from 0.8 mg/kg to 0.4 mg/kg based on the one-stage leaching tests of the carbonated stainless-steel slag. Ya-jun Wang et al. (Wang et al., 2020) performed the static thin-film accelerated carbonation and sequential leaching tests on the AOD steel slag and stated that the leachability of Cr increased with the growth of the carbonation ratio, reaching the highest leaching value of 7.9 mg/kg at 52% of the carbonation ratio. However, there was a dearth of further study on the influence mechanisms of the CO₂ accelerated treatment processes on the leachability of hazardous elements, which is important to prevent the potential of toxicity and promote the high blended proportion use of steel slag in the construction materials.

Herein, the influences of two different CO₂ treatment processes on the metal leachability and valence states were investigated based on the pastes of 80% steel slags. Two carbonation approaches were adopted for comparison, including hot-stage carbonation (applied during pretreatment of steel slag with simulated gas) and post carbonation (applied during post-curing). The Tessier sequential extraction was carried out to determine the different chemical profiles of the metal elements in these specimens. The mineralogical, CO₂ uptake, and microstructural properties were specifically investigated and compared to analyze the influence of carbonation on the leachability of metal elements in the end products.

MATERIALS AND METHODS

Materials
The original electric arc furnace steel slags (SS for short) were obtained from one steel plant in Shandong Province, China. The EAFs were ground and sieved to less than 200 μm, which is shown in Figure 1A. According to the results of Ca-Looping (CaL) process (Miranda-Pizarro, et al., 2016; Tian, et al., 2014) and fast absorption of CO₂ by steel slag (Dong, 2008; Stolaroff, et al., 2005), the hot-stage carbonation of SS was proposed to carbonate the free calcium oxide in SS. The carbonated EAF slags (CSS for short) were obtained by the raw EAF slags heated at 650°C in a 99.9 vol%CO₂ stream atmosphere under the atmospheric pressure, which is shown in Figure 1B. The reference cement (RC for short) was the 42.5-grade cement provided by Fushun Cement Co. Ltd. in Liaoning Province. Table 1 lists the chemical composition of SS, CSS, and RC determined by XRF. XRD and TG-DSC analysis of original EAF steel slag and carbonated EAF slag are shown in Figure 2. The results indicated that the most relevant mineralogical phases, including dicalcium silicate (Ca₂SiO₄), the RO phase, srebrodolskite (Ca₂Fe₂O₅), and Brownmillerite (Ca₂(Al,Fe)₂O₅) stayed almost the same after hot-stage carbonation, confirming the results of other studies on EAF slag (Wu et al., 2015). The carbonation efficiency (hereinafter referred to as EOC) was widely used to estimate the carbonation performance of various building materials or solid wastes by TG-DSC analysis (Chang and Fang, 2015; Liu and Wang, 2018). Mass change of calcite decomposition of the carbonated EAF slag samples was 1.13%, and the EOCs of the EAF slag samples were 3.45% after hot-stage carbonation process.

Production and Carbonation of Granulated EAF Slag Binders
Three different carbonation procedures were designed to investigate the effects of CO₂ on the performances of the produced EAF slag binders, as listed in Table 2. As indicated
TABLE 1 | Chemical composition of steel slag and reference cement.

| Sample                  | f- CaO | f- MgO | Fe₂O₃ | CaO  | SiO₂  | MgO  | Al₂O₃ | MnO  | P₂O₅ | TiO₂ | Cr₂O₃ | SO₃  | V₂O₅ | Na₂O  | ZnO  |
|-------------------------|--------|--------|--------|------|-------|------|-------|------|------|------|-------|------|------|-------|------|
| Raw EAF slag (R)        | 1.72   | 0.29   | 34.12  | 30.99| 16.3  | 7.76 | 3.78  | 3.3  | 1.49 | 0.7  | 0.44  | 0.277| 0.273| 0.258 | 0.025|
| Carbonated EAF slag (C) | 0.36   | 0.077  | 34.12  | 30.74| 16.58 | 7.63 | 3.92  | 3.3  | 1.5  | 0.696| 0.437 | 0.279| 0.273| 0.219 | 0.029|
| References cement       | 0.88   | —      | 3.62   | 63.56| 20.12 | 2.07 | 5.12  | —    | —    | —    | —    | —    | —    | —    | —    |

FIGURE 2 | XRD (A) and TG-DSC (B) analysis of original EAF steel slag and carbonated EAF slag. (1–6 of Figure 2A denote lamite, calcium silicate, srebrodolskite, iron silicon oxide, RO phase, and brownmillerite, respectively.)
in Table 2, C2C represents the carbonation applied both during the pretreatment of EAF slags and after the demoulding of steel slag paste binders, A2C indicates the carbonation applied only the post-curing treatment of the demoulding paste binders, and A2A refers to the control samples that were pretreated and cured without carbonation. The hot-stage CO2 pretreatment was applied during the pretreatment of EAF slags (SS) by importing the high purity CO2 gas into the tube reactor (T = 650°C, 99.9 vol% CO2 gas stream, and the atmospheric pressure), producing the carbonated EAF slags (CSS). The blend consisted of 80% SS/CSS and 20% RC was obtained by premixing the raw materials in a tumbling mixer. The water, and blend used to fabricate a paste specimen were mixed at a constant water-to-blend ratio of 0.4:1. The specimens were cast as follows: the dry-mixing of the blend were mixed for 5 minutes to ensure uniformity in the dry mix, after which water was added to the mixture, which was then mixed for a further 5 minutes. The paste binders were then cast into a 30 mm × 30 mm × 30 mm mold, prepared for the mineralogical and microstructural characterization. The pastes were demolded after curing in a standard curing chamber (T = 20 ± 2°C, RH 90%) for 24 h; and then A2A paste specimens transferred back into the curing chamber until 28 days curing age, the A2C paste and C2C paste specimens were placed immediately into a sealed carbonation chamber of 99.9 vol% CO2 concentration, 0.1 MPa CO2 pressure, and a temperature of 25 ± 2°C for 24 h, respectively. The pastes were broken into pieces with an iron hammer and immersed in absolute ethyl alcohol for 24 h to stop the further reaction. The pieces for microstructural analysis were grounded into powders with an agate mortar. The pieces for microstructural analysis were dried in a vacuum oven with silica gel at 50°C for 24 h.

**Testing and Characterization**

XRF analyses of the reference cement and steel slags were conducted on the equipment of ARLAdvant’X Intellipower XRD equipment with Cu Ka (λ = 0.15406 nm) ray from 10° to 70°. The Fourier transform infrared spectra (FTIR) were recorded on an infrared spectrometer (Nicolet IS5). The morphologies and microstructure of the binders were observed by scanning electron microscopy (SEM) coupled with energy dispersive spectroscopy (EDS) (Hitachi S-4,800). The pore size distribution were analyzed by a BET analyzer (Micromeritics ASAP2460). The Mastersizer 2,000 laser granularity meter was used to measure the particle size distribution. The surface atomic concentration and chemical states of the obtained samples were determined by XPS spectrometer (Thermo Fisher K-alpha). The procedure of Tessier sequential extraction was that the sequential

### Table 2: Carbonation procedures

| Notation | Carbonation procedure | Proportion/% | Note |
|----------|-----------------------|--------------|------|
| C2C      | hot-stage CO2 pretreatment and accelerated carbonation curing | 80% Carbonated EAF slag + 20% RC | for the binders that were made of carbonated EAF slag and demoulded into the post-curing treatment |
| A2C      | accelerated carbonation curing | 80% raw EAF slag + 20% RC | for the binders that were made of raw EAF slags and demoulded into the post-curing treatment |
| A2A      | no carbonation        | 80% raw EAF slag + 20% RC | for the control binders that were made of raw EAF slags and cured without carbonation |

### Table 3: Modified Tessier sequential extraction methods (Tan et al., 1997; Wei et al., 2010).

| Step | Speciation | Sequential extraction condition |
|------|------------|--------------------------------|
| I    | Exchangeable metals | 100 ml CH₃COONa solution (1 mol/L, pH = 8.2). Shaking extraction was continued for 1 h at ambient temperature |
| II   | Metals bound to carbonates | 100 ml CH₃COONa solution (1 mol/L, pH was adjusted to 5.0 with CH₃COOH). Shaking extraction was continued for 4 h at ambient temperature |
| III  | Metals bound to Fe/Mn oxides | Shaking extraction was continued for 5 h at 96°C with 100 ml NH₂OH solution (0.04 mol/L), volume fraction (25%) |
| IV   | Metals bound to sulfide compounds | 15 ml 0.02 mol/L HNO₃ solution and 25 ml 30%H₂O₂ (pH was adjusted to 2.0 with HNO₃ solution). Shaking extraction was continued for 3 h at 85°C, 25 ml NH₄OH-HNO₃ solution (3.2 mol/L, 20%) was added to the cooled solution and diluted to 100 ml by deionized water. Shaking was continued for 30min. |
| V    | Residue    | Digested by HNO₃+HCl+H₂O₂+HF (1:3:1:3) with a microwave |
leaching characteristics of the metal combined in the paste binders were conducted according to the methods (Tan et al., 1997; Bacon and Davidson, 2008) shown in Table 3. The metal concentrations of the leaching solution were tested by the ICP-OES (inductively coupled plasma optical emission spectrometer, Optima 7000DV).

RESULTS AND DISCUSSION
Sequential Leachability
The fraction of the trace elements (chromium, zinc, vanadium, and titanium) in the sequential leachates is presented in Figure 3. The accumulated active phase fraction of chromium (the sum of exchangeable chromium and chromium combined with carbonates) in the A2A, A2C, and C2C samples was 8.79, 7.62, and 2.06%, respectively. Meanwhile, the accumulated inactive phase fraction of chromium (the sum of chromium combined with Fe/Mn oxides, sulfide compounds, and the residue) in the A2A, A2C, and C2C samples were 91.21, 92.38, and 97.94%, respectively. Cr of the EAF slag paste binders gradually transformed into the inactive phases with the increase of the
The chromium oxides in the EAF slags existed in the form of non-hydrolyzable spinel phases or hydrolyzable silicate phases (Alejandro et al., 2011), and the zinc mineral compounds existed in zinc oxide and zinc ferrite (franklinite, ZnFe$_2$O$_4$) (Machado et al., 2006; Chen et al., 2011). The leaching of chromium, and zinc from in the Cr-bearing spinel phases, Mg-bearing spinel phases, and zinc ferrite phase were considered to be limited or even negligible (Zhao et al., 2019). Theoretically, the phases MgCr$_2$O$_4$, and ZnFe$_2$O$_4$ were carbonated by one static carbonation process. According to the thermodynamic data of the carbonatable mineral phases and the van’t Hoff equation, $\Delta G^\theta$ is the standard Gibbs free energy change, to determine the effect of the reaction factors and to judge the direction of the reaction. Based on the equilibrium equation of CO$_2$ and the mineral phases, the $\Delta G^\theta$ of the carbonation of MgCr$_2$O$_4$ ($-22$ kJ/mol) were less than that of Ca$_2$SiO$_4$ ($-116$ kJ/mol), CaO ($-135$ kJ/mol), and MgO ($-61$ kJ/mol) at $298$ K (Wang et al., 2020), and the $\Delta G^\theta$ of the carbonation of ZnFe$_2$O$_4$ ($-14$ kJ/mol) were also less, which indicated that these phases were stable and carbonated just to a less extent. The carbon efficiency was increased to that of C2C, which indicated that some parts of MgCr$_2$O$_4$ and ZnFe$_2$O$_4$ were involved in the carbonation. The separated Cr$_2$O$_3$ and Fe$_2$O$_3$ were hydrolyzed as one primary phase during the leaching (Li and Tsai, 1993).

**TG-DSC Analysis**

The TG-DSC curves of the uncarbonated area samples from A2A, carbonated area samples from A2C, and C2C after 1 day of curing age are shown in Figure 4. There were weight loss steps ranging from $25^\circ$C to $200^\circ$C, from $420^\circ$C to $460^\circ$C, and from $400^\circ$C to $750^\circ$C in the A2A curve, which corresponded to the removal of bound water and physically absorbed water, the decomposition of Ca(OH)$_2$, and the decomposition of calcite (Wang et al., 2018). The A2C and C2C samples showed one remarkable weight loss step ranging from $400^\circ$C to $950^\circ$C, corresponding to the breakdown of amorphous calcite (from $400^\circ$C to $700^\circ$C) and crystallized calcite (from $700^\circ$C to $950^\circ$C) (Jeong et al., 2017). The carbonation efficiency (hereinafter referred to as EOC) was widely used to estimate the carbonation performance of various building materials or solid wastes (Chang and Fang, 2015; Liu and Wang, 2018). The EOCs of the A2A, A2C, and C2C samples were $3.3$, $52$, and $67%$, respectively. The maximum carbonation efficiency of the binders was achieved by the simultaneous application of hot-stage CO$_2$ pretreatment and accelerated carbonation curing routes.

**XRD Patterns**

The XRD patterns of the three samples are shown in Figure 5. The A2A-28 days sample mainly consisted of the hydration product of the calcium silicates, such as some amorphous C-S-H gels, that cannot be detected though XRD; all of hydration products are prone to take part in the carbonation reaction and mainly produce carbonation products, such as calcite and monocarbonate (Wang et al., 2019). There were diffraction peaks of calcite were in the XRD patterns when the A2 paste was subjected to the accelerated carbonation curing treatment (Yu and Wang, 2011). The peaks of calcite were more prominent.
when the steel slag was used instead of the A2 EAF slag mixture under hot-stage CO$_2$ pretreatment and the paste was put into the accelerated carbonation. These results suggested that Ca-bearing minerals were carbonated, which corresponded to the results of the phenolphthalein indicator testing and thermodynamic calculation (Wang et al., 2019). The calcite existed in the crystalline and amorphous forms, and the content of CaCO$_3$ cannot be obtained by XRD analysis, which was demonstrated by the TG-DSC study.

**XPS Analysis**

XPS analysis results are shown in **Figure 6** and **Table 4** to demonstrate the chemical state of the elements in the A2A-28d, A2C-1d, and C2C-1d test samples. The Si 2p peaks of the A2A-28d, A2C-1d, and C2C-1d test samples were respectively located at 99.89, 100.95, and 104.43 eV, which indicated that the binding energy of Si 2p increased after the carbonation treatments. The loss of non-bridging oxygens (Si–O–Ca moieties etc.) tended to increase the Si 2p binding energy and related silicate polymerization (Black et al., 2003b). The carbonation reaction increased the consumption of the calcium, leading to a decrease in the Ca/Si ratio, and ultimately forming a Q4 silicate (Black et al., 2007) and calcite (Richardson, 1999). The researchers (Black et al., 2003a; Black et al., 2006; Okada et al., 1998) found that the binding energy of Si had contributed to that of silicate tetrahedral polymerization. The researchers (Black et al., 2004) discovered that there was a strong negative correlation between the Si binding energy and Ca/Si ratio in the Ca-bearing and Si-bearing hydrates phase. The Si 2p peaks of the A2C and C2C samples shifted to higher binding energy, which was slightly broader than that of A2A and consistent with the EOC trend. The broadening of the peak indicated the silicate structure with a higher disorder. There were Ca 2p3/2 and Ca 2p1/2 peaks in the Ca 2p XPS spectra. The changing trend of the Ca binding energy peaks resembled that of Si, and the centre order of the Ca binding energy peak was roughly determined as C2C > A2C > A2A, which suggested that the enhancement of carbonation reaction led to the increase of the Ca binding energy. The C2C sample had much broader Ca 2p binding energy peaks than A2C and A2A, which was due to the differences in the diversity of the products (monocarbonate, calcium carbonates, or Ca and Si-bearing hydrates of the C2C sample). The C 1s peaks of the A2A-28d, A2C-1d, and C2C-1d test samples at lower binding energy were respectively located at 283.10, 282.98, and 286.48 eV, which was assigned to as “adventitious carbon” (Black et al., 2008). The C 1s peaks at higher binding energy (only at 287.88 and 291.68 eV in the A2C-1d and C2C-1d samples) were attributed to carbonate species by combining with the Ca 2p spectra. The carbonation

**TABLE 4 | XPS fitting analysis of valence of Cr ions at A2A-28d, A2C-1d, and C2C-1d samples.**

| Sample     | Area(P) CPS.eV (Cr$^{2+}$) | Area(P) CPS.eV (Cr$^{3+}$) | Area(P) CPS.eV (Cr$^{6+}$) | Atom% (Cr$^{2+}$) | Atom% (Cr$^{3+}$) | Atom% (Cr$^{6+}$) |
|------------|---------------------------|---------------------------|---------------------------|-------------------|-------------------|-------------------|
| A2A-28d    | 68.25                     | 48.39                     | 32                        | 45.77             | 32.61             | 21.63             |
| A2C-1d     | —                         | 66.84                     | 146.03                    | —                 | 31.29             | 68.71             |
| C2C-1d     | —                         | 52.54                     | 34.83                     | —                 | 60.05             | 39.96             |

**FIGURE 7 | FTIR spectra of the uncarbonated area sample from A2A-28d, carbonated area samples from A2C-1d, and C2C-1d.**
of the A2C-1d and C2C-1d samples was slight, and there was one carbonate peak just about visible in the XPS spectra (Gonzalez-Elope et al., 1990). There were binding energy peaks of Cr$^{2+}$, Cr$^{3+}$, and Cr$^{6+}$ in the Cr 2p XPS spectra, and the valence analysis results are listed in Table 4. According to the literature (Wang et al., 2015), the steel slag had the characteristic peak of Cr, such as Cr$^{2+}$, Cr$^{3+}$, and Cr$^{6+}$, and the valence states of Cr$^{2+}$ and Cr$^{3+}$ were the main occurrence forms. When pH $\geq$ 11, most of Cr$^{6+}$ exists in the form of CrO$_4^{2-}$; when pH < 1.2, most of them exist in the form of Cr$_2$O$_7^{2-}$; under alkaline conditions, Cr$^{3+}$ mainly exists in the form of Cr(OH)$_3$, and under acidic conditions, Cr$^{3+}$ exists in the form of ions (Wei et al., 2010). Cr$^{6+}$ partly transformed into Cr$^{3+}$ due to the influence of other elements and partial pressure during the treatment (Silva and Monteiro, 2007). The changing trend of the Cr 2p binding energy peaks resembled that of Si 2p after carbonation treatment, the fraction of Cr$^{3+}$ in the C2C sample was far greater than that of the A2A sample, which indicated that carbonate formation, mineralogical changes and the decrease of pH (lower pH) promote the leaching of one oxidation state of Cr or different leaching rates of Cr valences, corresponding to other research results (Wu et al., 2016). XPS is a general semi-quantitative analysis method for analysing the valence states of Cr element, only reflecting the surface atomic characteristics. The valence distribution of Cr can be studied by simplifying the composition of the slag and other analysis methods.

**FTIR Analysis**

Figure 7 shows the FT-IR spectra of the uncarbonated area in the A2A-28d sample and carbonated area in the A2C-1d and C2C-1d samples. There were asymmetric stretching ($\nu_3$) vibration peaks of [CO$_3$]$^{2-}$ at 1,424.25, 1,427.5, and 1,448.86 cm$^{-1}$ in these samples (Chang et al., 2019). The out-of-plane bending ($\nu_4$) peaks of the C-O bond in calcite (Vagenas et al., 2003) were located at 875.14, 873.19, and 873.52 cm$^{-1}$. The in-plane bending peaks ($\nu_3$) of the C-O bond in the aragonite and calcite (Chang et al., 2019) were centered at 714.01 and 711.6 cm$^{-1}$. The carbonation promoted the formation of crystalline and amorphous calcite. The asymmetric stretching peaks ($\nu_3$) of
the Si-O bond in the silica gel (Ukwattage et al., 2015) were located at 989.94, 1,042.09, and 1,043.34 cm$^{-1}$, and the out-of-plane bending ($\upsilon_4$) of the Si-O bond (Moon and Choi, 2019) were detected at 516.10, 568.7, and 576.29 cm$^{-1}$, which indicated the formation of silica gel during carbonation. The peaks at 3,640–3,645 cm$^{-1}$ were assigned to the stretching vibration of the O-H bond, while the peaks at 3,420–3,450 cm$^{-1}$ were contributed to the vibration of the H$_2$O bond in the hydration products (Lukas, 1976). Compared with the uncarbonated sample of A2A, the bands of the carbonated samples at 870–880 cm$^{-1}$ decreased, but the bands at 510–580, 700–750, 1,000–1,100, and 1,400–1,500 cm$^{-1}$ increased, which represented that the calcium silicates gradually reacted with CO$_2$ to form calcites and silica gels during the carbonation.

**SEM-EDS Analysis**

Figure 8 shows SEM-EDS images of these samples. According to the A2A and A2C images, the iron-containing substances (such as RO phase and brownmillerite) were involved in the carbonation reaction just to a less extent, and the uncarbonated calcium silicates core surrounded the Ca-leached and Si-rich layers (Quaghebeur et al., 2015). Physical encapsulation of the carbonate, which generated under such a carbonation ratio, reduced the hydrolysis of the primary mineral phases and then immobilized the metal, such as chromium (Spanka et al., 2016).

**BET Analysis**

The porosity volume, and average pore size of the original slag, carbonated slag, A2A-28d, A2C-1d, and C2C-1d steel slag-based binders are shown in Figure 9. By comparing the pore size distribution curves of the samples (as shown in Figure 9), the proportion of 2.5–10 nm mesopores of sample A2C and C2C is higher than that of sample A2A, which would result in higher performance. The average pore sizes of the three samples were 17.44, 7.22, and 6.77 nm, respectively. The pore size was reduced in the blocks, which was caused by the newly formed calcites and silica gels covering the pores of the matrix and precipitates joining together (Pizzol et al., 2014; Muller and Scrivener, 2017). The above findings fully vindicated that both carbonation processes reduced the pore area and porosity of the reaction matrix, thus promoting strength.

**CONCLUSION**

In conclusion, the carbonation processes of the steel slag binders were investigated, and the relationship between the carbon efficiency of the carbonation processes and the metal leachability as well as valence states minerals changes. The synergistic use of hot-stage CO$_2$ pretreatment and accelerated carbonation curing obtained the maximum CO$_2$ storage of 52% based on the binders of 80% steel slags. The hot-stage carbonation treatment facilitated particle agglomeration, minerals remodeling, and calcite formation. The carbonation curing of the steel slag paste resulted in the formation of the amorphous CaCO$_3$, calcite crystalline and Si-bearing hydrates that covered the pores of the matrix, and silicate structure with a higher disorder. The C2C binder had the lowest leachability of chromium, vanadium, and titanium, and the A2C binder had the lowest leachability of zinc. The exchangeable leachability of chromium was 1.99 mg/kg in the A2A binder, which declined to below the detection limit in the A2C sample and C2C binder after two carbonation processes. The comprehensive utilization of hot-stage carbonation and accelerated carbonation curing methods prevented the leaching of harmful metals, which provided one
reference on the reuse of steel slags as the raw building materials in one higher and more friendly way.

**DATA AVAILABILITY STATEMENT**

The original contributions presented in the study are included in the article/Supplementary Material, further inquiries can be directed to the corresponding author.

**AUTHOR CONTRIBUTIONS**

YL: Conceptualization; Methodology; Investigation; Data curation; Formal analysis; writing-original draft. JF: Funding acquisition; Writing—Review Editing. SL: Writing—Review Editing. XA: Conceptualization. YK: Investigation. LW: Investigation; Supervision.

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