Ocean Acidification Mitigates the Negative Effects of Increased Sea Temperatures on the Biomineralization and Crystalline Ultrastructure of Mytilus

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Negative impacts of global climate change are predicted for a range of taxa. Projections predict marked increases in sea surface temperatures and ocean acidification (OA), arguably placing calcifying organisms at most risk. While detrimental impacts of environmental change on the growth and ultrastructure of bivalve mollusk shells have been shown, rapid and diel fluctuations in pH typical of coastal systems are often not considered. Mytilus edulis, an economically important marine calcifier vulnerable to climate change, were exposed to current and future OA (380 and 1000 ppm pCO2), warming (17 and 20 °C), and ocean acidification and warming (OAW) scenarios in a seawater system incorporating natural fluctuations in pH. Both macroscopic morphometrics (length, width, height, volume) and microscopic changes in the crystalline structure of shells (ultrastructure) using electron backscatter diffraction (EBSD) were measured over time. Increases in seawater temperature and OAW scenarios led to increased and decreased shell growth respectively and on marginal changes in cavity volumes. Shell crystal matrices became disordered shifting toward preferred alignment under elevated temperatures indicating restricted growth, whereas Mytilus grown under OAW scenarios maintained single crystal fabrics suggesting OA may ameliorate some of the negative consequences of temperature increases. However, both elevated temperature and OAW led to significant increases in crystal size (grain area and diameter) and misorientation frequencies, suggesting a propensity toward increased shell brittleness. Results suggest adult Mytilus may become more susceptible to biological determinants of survival in the future, altering ecosystem structure and functioning.

Keywords: multiple stressors, climate change, biomineralization, mussels, environmental variability, functioning

INTRODUCTION

Over the last century, atmospheric concentrations of CO2 have increased at an unprecedented rate resulting in changing environmental conditions on both land and sea. In the ocean, the increased absorption of CO2 due to higher atmospheric partial pressures is altering chemical reactions and driving declines in pH and (CO32−) (Doney et al., 2009) – a phenomenon referred to as ocean
acidification (OA). At current CO₂ emission rates, by 2100 atmospheric CO₂ concentrations are predicted to exceed 1000 ppm in some instances (Stocker et al., 2013), leading to reductions in pH of between 0.3 and 0.4 units from today’s conditions (Doney et al., 2009). For marine life, changes in pH can have significant, detrimental effects on their structure and functioning (Lemasson et al., 2017a, 2018) including abnormal larval development (Kurihara et al., 2007; Kurihara, 2008), increased pressure on acid-base regulatory mechanisms (Lindinger et al., 1984; Thomsen and Melzner, 2010; Scanes et al., 2017) and changes to behavior (Queiros et al., 2015; Sadler et al., 2018).

For calcifying species, OA can be especially problematic due to disruption of the biomineralization process that is so crucial to the development of shells, exoskeletons and tests. In bivalve mollusks, the shell often consists of two polymorphs of calcium carbonate: the outer prismatic layer comprising calcite crystals, and the inner nacre layer comprising aragonite crystal tablets (Gazeau et al., 2013). Both polymorphs are formed within the extrapallial space following the catalytic conversion of CO₂ to bicarbonate by the enzyme carbonic anhydrase (Marin and Luquet, 2004). This catalysis is a crucial process for calcium carbonate crystal nucleation, growth and orientation (Nakahara, 1991; Choi and Kim, 2000; Olson et al., 2013). Under normal conditions, these crystals are arranged as horizontal sheets (Checa et al., 2006; Hahn et al., 2012) but under predicted future climate conditions, crystal formation can become disordered and the crystals themselves misoriented or porous (Hahn et al., 2012; Fitzet et al., 2014a,b; Li et al., 2015; Meng et al., 2019; Figure 1). It is argued that these changes will have negative consequences for shell integrity and material properties (e.g., Beniash et al., 2010; Fitzet et al., 2015b), impacting the potential for mollusks to withstand physical and biological stress (Sadler et al., 2018).

The effects of OA on mollusk shell calcification and ultrastructure have largely been considered under fixed pH conditions; the experimental pH treatments are tightly controlled within a small pre-defined range and considerable negative consequences for organisms have been shown (e.g., Dupont et al., 2013). Yet increasingly, in situ measurements of pH in a range of coastal marine habitats including estuaries, kelp forests, coral reefs, and upwelling regions (Hofmann et al., 2011) reveal considerable natural fluctuations in pH (of up to 0.8 units) over short (diel) timescales (e.g., Lemasson et al., 2018). Where studies have incorporated pH fluctuations within their experiments either experimentally or by sampling organisms from sites with varying degrees of environmental fluctuation (Dufault et al., 2012; Comeau et al., 2014; Frieder et al., 2014; Ramajo et al., 2019), the impacts of OA have been less severe, perhaps due to transient exposure to less acidified conditions (Wahl et al., 2016) or that organisms originating from sites with naturally higher fluctuations in pH may be more tolerant or locally adapted to OA (Pansch et al., 2014).

Increasingly, studies are beginning to include the potential interactive effects of temperature in order to better simulate the ocean acidification and warming (OAW) environmental scenarios predicted for the end of the century. These studies have largely focused on macro-morphological change but often results appear contradictory or indicate a degree of context specificity (sensu Pansch et al., 2014); differences that may be driven by location, species, or life-history stage. For example, Lagos et al. (2016) found elevated temperatures mitigate negative impacts of OA on shell growth in *Argopecten purpuratus*. In *Littorina littorea*, Melatunan et al. (2013) found elevated temperature or reduced pH were equally detrimental to growth but effects increased in severity when combined, and Lemasson et al. (2018) found elevated temperature but not OA affected growth in oysters.

Relatively few studies have assessed the combined effects of OAW scenarios on shell ultrastructure. Independently, temperature has been shown to induce significant crystal misorientation (Olson et al., 2013), changes in thickness (Olson and Gilbert, 2012; Gilbert et al., 2017) and alteration in the overall ultrastructure (Füllenbach et al., 2014). These studies, however, have mainly been in the context of paleo-environmental proxies which arguably have limited relevance to the current, rapid changes in temperature driven by anthropogenic activities. Of the studies that have investigated the combined effects of OA and warming (e.g., Fitzet et al., 2014a, 2015b; Li et al., 2015), the evidence suggests that elevated temperatures can exacerbate ultrastructure disruption caused by elevated pCO₂, although the chosen experimental conditions are not necessarily those marine organisms are expected to face in the next century.

Understanding change in the morphological characteristics of shell ultrastructure, such as changes in crystal orientation, the occurrence of coincident site lattice (CSL) boundaries between crystals, and crystal size, may provide important insights in to how organisms will respond to future climate change. Misorientation – the difference in orientation between adjacent crystals – is expected to be relatively rare under low stress conditions as the organic matrix is able to maintain a well-ordered ultrastructure. Under environmental stress (e.g., OAW scenarios), that organic matrix may breakdown leading to disordered ultrastructure and higher instances of crystal misorientation (Olson et al., 2013). Crystal misorientation can lead to an increased frequency of CSL boundaries; the boundary where the positions of a proportion of lattice sites coincide at the boundary segments between two adjacent crystals (Fortes, 1972). Some material science studies have found that low angle grain boundaries and low Sigma (Σ)-value CSL boundaries can influence material properties with increases in hardness and resistance to the spread of intergranular cracks (Lehockey et al., 2004; Arafin and Szpunar, 2009). An increased frequency of certain CSL boundaries in biominalerized structures such as calcified mollusk shells, may be an important determinant of resistance to mechanical stresses like those experienced by bivalves under predation pressure from decapods and gastropods (Elner, 1978; Sadler et al., 2018). Changes in crystal size may also indicate impacts of environmental conditions on crystal nucleation and growth.
mechanisms. Such changes have been shown over paleo-
geological timescales (e.g., Olson and Gilbert, 2012; Gilbert 
et al., 2017) but have yet to be considered in the context 
of rapid environmental change that marine organisms are 
experiencing today.

This study investigated the extent to which the combination 
of elevated pCO$_2$ concentrations and temperature can alter 
the growth and ultrastructure of *Mytilus* sp. Using an 
experimental mesocosm system able to replicate future OAW 
scenarios that encompasses natural fluctuations in pH, we 
compare changes in (i) whole organism shell growth, and 
(ii) crystal orientation, size, and lattice boundary formations 
using cutting-edge electron backscatter diffraction techniques 
(EBSD).

**MATERIALS AND METHODS**

**Mussel Culture**

Mussels were collected from Queen Anne's Battery, Plymouth 
(50°36'40.82''N, 4°13'11.61''W), cleared of epibionts with a 
metal scraper, and transported back to the lab in buckets.
of seawater. Animals were placed in tanks containing UV-treated filtered seawater collected from the sampling location within 1 h of collection and allowed to acclimate to laboratory conditions (salinity = 29.7 ± 2.3; temperature = 17°C ± 0.20; pH 8.00 ± 0.04) for 2 weeks. After acclimation, 36 mussels were randomly allocated to 12 chambers (3 mussels per chamber) for 8-week. These chambers were filled with fresh UV-treated filtered sea water, also collected from the sampling location. Mean salinity was marginally higher (32.6 ± 1.3) than in the acclimation treatment as a result of natural fluctuations in salinity due to rainfall and riverine inputs. Each chamber represented one of four environmental scenarios (n = 3). By the end of the century, future climate change scenarios predict increases in sea surface temperature of 2–4°C and increases in atmospheric pCO2 of ∼600–800 ppm (IPCC, 2014) depending on the emissions (RCP) scenario. Here, we adopted a temperature increase of 3°C and pCO2 increase of 620 ppm, resulting in the following four combinations: (1) 17°C × 380 ppm pCO2; (2) 17°C × 1000 ppm pCO2; (3) 20°C × 380 ppm pCO2; and (4) 20°C × 1000 ppm pCO2. Temperature treatments were maintained using recirculating thermostatically controlled heated water baths (Aqua Medic, United Kingdom) bubbled with air or 1000 ppm CO2-enriched air (see Lemasson et al., 2018). CO2 concentrations were controlled using a CO2 gas analyzer (LI-COR LI-820, United States). Natural fluctuations in pH result from this approach as pCO2 concentrations remain static and pH changes occur as a result of fluctuations in atmospheric pressure, although these fluctuations in pH were relatively small with a maximum standard deviation of 0.1 pH units in any treatment (Supplementary Table S1). Seawater pH (SevenExcellence Multi-parameter pH probe), temperature (Omega HH802U temperature probe) and salinity (HI96822 Seawater Refractometer) were tested daily and total alkalinity (Omega HH802U temperature probe) and salinity (HI96822 Seawater Refractometer) were measured three times per week. Mussels were fed every other day with 10 mL of Shellfish Diet 1800 at a concentration of ∼4500 cell mL−1 (Thomsen and Melzner, 2010; Fitzer et al., 2014b). Four biometric parameters: total wet mass, shell length, shell width, and shell height were measured before and after an 8-week exposure period.

Body volume (cm3) of all individuals (N = 36) was calculated using the volume of water displaced (mL) by the mussel in a volumetric cylinder containing sea water (density of 1.025 g/cm3 at 20°C, salinity = 35). This approach controls for any water retained within the mantle. Displacement (mL) of an individual was measured at time zero (t0) and after 8-week (t8) of exposure to all experimental treatments. Sea water density was measured using a hand-held digital refractometer (D&D, The Aquarium Solution Ltd., United Kingdom). Shell dimensions (length, width, height) were measured using digital calipers (Mitutoyo, Japan) to the nearest 0.1 mm, also at t0 and t8 (Figure 2A).

Ultrastructure Data Sampling Procedure

After 8-week, all mussels were euthanized and the soft tissue removed. Two mussels per treatment were randomly chosen from the control (17°C × 380 ppm pCO2), elevated temperature (20°C × 380 ppm pCO2), and elevated temperature and pCO2 treatment (20°C × 1000 ppm pCO2) and a cross-section from the lip area of each mussel removed using a high-speed cutting wheel (Dremel, United States). Mussels from a control temperature and elevated pCO2 treatment (17°C × 1000 ppm pCO2) were not used for EBSD analysis due to logistical constraints and given that this scenario is not expected under future climate scenarios. Cross-sectional cuts were specifically made to allow examination of the ultrastructure between the prismatic and nacre layers (Figure 2B) along the growth direction in the most recently formed shell material (Checa et al., 2006; Hahn et al., 2012).

Shell samples were mounted in epoxy resin (Struers Epofix), ground [Buehler Carbimet Grit® 600 (P1200) pad] and sequentially polished with 6 µm then 1 µm diamond suspensions (Buehler MetaDi®) until a glossy appearance was achieved. Samples were then polished using a vibratory polisher (Buehler VibroMet²®) with a 0.05 µm polishing suspension (MasterPrepTM) diluted in distilled water. Samples were examined under an optical microscope (Meiji Techno EMZ-13TR, Infinity 1) to check the quality of the polished surface (Figure 1). Samples were preserved in ethanol before carbon sputter-coating (Quorum Technologies Q150T ES, United Kingdom).

Samples were examined in a variable-pressure scanning electron microscope (SEM) (JEOL 6610 VP-SEM, Japan) under high vacuum at 15 kV. Magnification, spot size, and working distance were modified between samples to optimize data acquisition. Crystallographic data were obtained using a Nordlys Nano EBSD detector with Aztec software, and analyzed using HKL CHANNEL5 post-processing software (Oxford Instruments, United Kingdom). Grains were detected using a minimum pixel number of 10 (default setting) and a critical misorientation angle of 3° as the defining criteria (Goetz et al., 2011).

Ultrastructure Parameters

Differences in the ultrastructure between samples were assessed using a combination of inverse pole figure (IPF) orientation maps, contoured pole figures, frequency distributions of grain boundary misorientation angles, Σ-value CSL boundaries, grain area, and grain diameter metrics. IPF maps using the z-direction (IPFz; perpendicular to the sample surface) as the sample reference direction were used to demonstrate patterns of variability in orientation in shell samples. IPF maps present patterns of crystallographic orientation using different colors to illustrate different orientations of EBSD measurement pixels relative to a fixed orientation. In a well-ordered ultrastructure, IPF maps would be expected to show similar orientations throughout much of the mapped site, or with different orientations restricted to definable regions. The majority of grains should appear to face a similar direction along the sample surface. In a more disordered ultrastructure, the mapped sites would be expected to show a wide variation in orientation with the direction that the grains face, relative to the sample surface, appearing more random.

Contoured pole figures (1 point per grain analysis) were used to demonstrate the variation in crystallographic orientation of individual grains relative to different planes. Ordered/orientated
Crystal lattices form clusters around specific orientations (i.e., single crystal fabrics), which indicates uninhibited growth, whereas disordered/misoriented crystal lattices demonstrate "preferred alignment" indicating inhibited growth. The probability of grain orientation angles (multiple uniform densities; MUD) are illustrated using pole figures.

The frequency distribution of grain boundary misorientation angles demonstrates variation in co-orientation between adjacent grains. Misorientation angle is defined as the rotation angle required to map the lattice of the boundary segment of a grain to the lattice of the adjacent boundary segment (Oxford Instruments, 2015) and the frequency distributions of the Σ values of CSL boundaries can demonstrate variation in co-orientation and fracture resistance between grains. The Σ values reflect the proportion of coincident lattice sites between the two boundary segments. For example, in Σ 3 boundaries, one out of every three lattice sites are coincident between the boundary segments (Fortes, 1972). IPFz maps were annotated with the positions of grain and CSL boundaries to demonstrate the distribution of different misorientation angle ranges and Σ values, respectively.

Grain size was used to test for changes in grain formation and growth. Both grain area and diameter were included as measures of grain size, and plotted as frequency distributions, to account for the potential effect of grain shape on the latter.

Three regions (subsets) were defined for each treatment to further explore the variability in ultrastructure parameters within a single shell sample. Each region was chosen to assess crystalline structure in the outer periostracum layer (subset 1), center (subset 2), and inner nacre layer (subset 3) to determine if different regions of the shell responded differently to environmental stresses (Figure 2B shows the transverse EBSD scan direction). Statistical tests were only conducted between subsets from the same sample.

**RESULTS**

**Mussel Volume and Growth**

There were significant changes in body volume depending on temperature and pCO₂ conditions after 8-week ($F_{1,32} = 11.9, p < 0.01$). Mussel body volume reduced under either elevated temperature (20°C) or pCO₂ (1000 ppm), but marginally increased under control (0.13 ± 0.09 mL) and elevated temperature and pCO₂ conditions (0.05 ± 0.02 mL) (Figure 3).

There were significant reductions in both shell width and length under the OAW scenario (20°C × 1000 ppm; width: $F_{1,32} = 9.2, p < 0.01$; length: $F_{1,32} = 5.6, p < 0.05$) of 1.05 ± 0.48 mm and 0.86 ± 0.44 mm, respectively. There was no change in morphometrics under control (17°C × 380 ppm), increased temperature or increased pCO₂ conditions (Figures 3B,C). There was no significant change in shell height among treatments ($F_{1,32} = 0.56, ns$, Figure 3D), although trends suggest a decrease in shell height especially...
FIGURE 3 | Change in (A) body volume (mL), (B) shell width (mm), (C) length (mm), and (D) height (mm) after 8-week exposure to four climate scenarios. Error bars show standard error. Results of post hoc pairwise comparisons are shown, where shared letters indicate no significant difference between pairwise groups (p > 0.05), and different letters indicate significant differences (p < 0.05) between pairwise groups.

under control conditions, albeit reduction among individuals being highly variable.

**Ultrastructure Parameters**

**Grain Area and Diameter**

There were significant differences in the frequency distributions of grain areas and diameters among all experimental treatments after 8-week (Kolmogorov–Smirnov paired comparisons; p < 0.001 for all comparisons). Both grain area and diameters tended toward smaller dimensions (left skew; Figure 4). Grain areas and diameters were smallest in the control treatment (17°C × 380 ppm), largest in the elevated temperature treatment (20°C × 380 ppm), and of intermediate size in the OAW (20°C × 1000 ppm) treatment. Median grain areas were 3.6 × and 2.2 × larger, and diameters 1.9 × and 1.5 × larger, in 20°C × 380 ppm and 20°C × 1000 ppm treatments respectively than in the control treatment (Figure 4). While median grain area was lower in the OAW scenario than the temperature only treatment, the maximum grain area observed in mussels from the OAW treatment was up to 7.8 × larger (12,887 µm² vs. 1680 µm²) and 16.5 × larger than the temperature only (20°C × 380 ppm) and control treatments, respectively (Figure 4).

**Crystallographic Orientation and Disorientation**

Change in crystallographic orientation and increased prevalence of misorientation angles were revealed by EBSD (Figure 5). Under control conditions, crystal orientation was largely well-ordered with clear evidence of single crystal fabrics (uninhibited growth) across all crystal planes (Figure 6). There were few deviations from the expected crystallographic orientation angle of 3–10° and a low prevalence of high misorientation angles (16% > 20°). Under elevated temperature (20°C × 380 ppm)
and OAW (20° × 1000 ppm) scenarios, crystals along the 0001 plane also showed propensity toward single crystal fabrics; the orientation of the crystals in the elevated temperature scenario were also rotated.

The 11-20 and 10-10 planes of the elevated temperature treatment exhibited a major shift toward preferred alignment and a higher prevalence (~23.1%) of misorientated angles (>20°) at the grain boundaries (Kolmogorov–Smirnov tests, p < 0.01; Figures 5, 6). This shift toward larger misorientation angles (>20°) was even more apparent in the OAW-treated mussels (41%), although crystals appeared to continue to grow as a single crystal fabric in all planes (Figure 6).

Comparison of growth between the external-facing (prismatic) layer and internal-facing (nacreous) revealed differing responses to changed environmental conditions. Focusing on the 11-20 plane only (although a similar pattern was also apparent in the 10-10 plane), *Mytilus* reared in elevated temperatures only (20°C × 380 ppm) showed evidence of restricted growth (preferred alignment) across the entire transverse section of the shell, in particular, within the nacreous region (Figure 7H). In comparison, *Mytilus* sp. grown under control (Figures 7A,D,G) and OAW conditions continued to demonstrate single crystal fabric (uninhibited) growth, if not, with marginally lower “intensity” (Figures 7C,F,I).

**Coincident Site Lattice Boundaries**

Coincident site lattice boundaries were most frequent in the elevated temperature treatment (20°C × 380 ppm) followed by the OAW scenario and then the control treatment (Figures 5B,D,F). CSL boundaries in the OAW and control samples were restricted to isolated patches, whereas under elevated temperature only, CSL boundaries were distributed throughout the region, although there was no significant difference in the occurrence of these boundaries among treatments.

**DISCUSSION**

By the end of the century, future climate change scenarios predict increases in sea surface temperature of ~4°C and increases in atmospheric pCO₂ of ~600–800 ppm (IPCC, 2014). Exposure of the commercially valuable mussel species, *Mytilus*, to conditions predicted for 2100 led to changes in body volume and shell growth at the macroscale, and alteration of shell crystallization (ultrastructure) at the microscale, although negative consequences mostly occurred under temperature as a single stressor rather than in combination with increases in pCO₂. Results suggest an antagonistic relationship between temperature and pCO₂ stressors on biological functioning of *Mytilus*.

Body volume was reduced under elevated temperature or elevated pCO₂ treatments, but marginally increased under control conditions or when both temperature and pCO₂ were elevated. In contrast, shell length and width reduced under elevated temperature and pCO₂ (20°C × 1000 ppm) conditions, but marginally increased under elevated temperatures (20°C) only. Body volume can act as a proxy measure for the condition of the mussel; change in the mass of soft tissue filling the mantle alters the shell: tissue: mantle-volume ratio and thus water displacement, thereby indicating change in energy utilization over time in somatic tissues development.

**FIGURE 4** Distribution of grain areas (µm²) and diameters (µm) within regional subsets of *Mytilus* sp. exposed to three temperature (°C) and pCO₂ climate scenarios for 8-week. Thick lines of boxplots represent the median value, hinge lengths (end of box) represent the 25 and 75% quartiles from the median, and whiskers represent the 1.5 times the interquartile range (IQR) beyond the hinge. Outliers are shown as black circles.
or maintenance (e.g., Melzner et al., 2011; Huning et al., 2013). The reductions in volume associated with elevated water temperature or pCO$_2$ might initially suggest an increased metabolic cost to individuals preventing soft issue maintenance, but concurrent increases in shell width and length instead suggest a reallocation of energy toward shell maintenance.

![Image of Figure 5](image-url)

**FIGURE 5** | Inverse pole figure (IPF) maps with grain boundaries (left column) and coincident site lattice (CSL; right column) boundaries of Mytilus sp. shell samples grown under three climate scenarios: (A,B) $17^\circ \text{C} \times 380$ ppm (Control); (C,D) $20^\circ \text{C} \times 380$ ppm (Ocean warming without acidification); and (E,F) $20^\circ \text{C} \times 1000$ ppm (Ocean warming with acidification). Bottom: Crystal orientation color inset key; grain boundary misorientation (strain) angles; and CSL Sigma boundaries (right column). All plots are orientated in the same way i.e., the left-hand side of the image is the outer-facing periostracum surface and the right-hand side is the inner-facing (nacre) surface (as illustrated on panel B).
under these single stressor conditions. Previous studies have shown organisms can induce shell thickening (Bibby et al., 2007; Melatunan et al., 2013) or repair damage to their shells (Coleman et al., 2014) through reallocation of energy to specific physiological processes, although OAW scenarios can alter this response. For instance, Li et al. (2016) showed that OAW implemented a compensatory acid-base mechanism, metabolic depression, and up-regulation of physiological responses leading to change in calcification rates, the calcium and carbon content within shells, as well as change in shell ultrastructure. Here, the notable decreases in shell length and width under the elevated temperature and pCO$_2$ (20°C × 1000 ppm) scenario support these previous findings, suggesting depression of some energetically expensive processes (i.e., soft tissue maintenance) and up-regulation of other physiological and biochemical functions. The observed increases in shell length and width under elevated pCO$_2$ and ambient temperature conditions (17°C × 1000 ppm) also suggests that mussels may have some capacity to up-regulate biomineralization to maintain or grow shell material under OA (Bibby et al., 2007; Melatunan et al., 2013; Coleman et al., 2014, but see Leung et al., 2020); a process that may be driven by an up-regulation of ion and proton transport genes associated with organic matrix formation (Glazier et al., 2020).

There were also changes in the ultrastructure of mussels grown under future climate conditions. Under elevated temperature, the shell ultrastructure became disordered shifting from single crystal fabrics toward preferred alignment suggesting "restricted"
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FIGURE 7 | Contoured pole figures (1 point per grain analysis) showing variation in crystallographic orientation along the 11-20 crystal plane in three regions (prismatic; center; and nacreous layers) of the ultrastructure of Mytilus sp. grown for 8-week under three climate scenarios: Control (A,D,C); Elevated Temperature (B,E,H); and Elevated Temperature and pCO$_2$ (C,F,I). The MUD value for the three treatments varies from 98.64 (baseline, 17 × 380) to 154.81 (subset 1, 20 × 1000).

growth, whilst area and diameter of individual grains became significantly larger. In contrast, under elevated temperature and pCO$_2$, crystalline structures remained ordered as a single crystal fabric, in the same way as control mussels, suggesting OA conditions may be mitigating the effect of temperature. Both future scenarios led to increased frequencies of crystal misorientation; an indicator of stress that signifies the formation of microcracks and weakening of the shell. Previous studies have shown bivalves grown under increasing pCO$_2$ produce a stiffer, harder calcite (prismatic) layer that is more brittle, whereas the aragonite (nacre) layer may become less stiff and softer (Beniash et al., 2010; Fitzet et al., 2015a) but see Hiebenthal et al. (2013), Mackenzie et al. (2014). While not tested here, the elongation of the crystals grown under elevated temperature and OA (20°C × 1000 ppm) suggests greater crystal malleability, not just within the nacre region of the shell but across the entire transverse section of the shell. Our results are therefore, in part, contradictory to previous studies that have reported disordered shell ultrastructure under elevated pCO$_2$ in Mytilus (Hahn et al., 2012; Fitzet et al., 2014a,b, 2016) and other mollusks (Beniash et al., 2010); an effect argued to be further exacerbated under elevated temperature (Fitzer et al., 2014b). It should be noted, however, that our experimental future climate temperatures (20°C) were considerably higher than those used in Fitzet et al. (2014b).

Change in the physiological mechanisms that underpins ultrastructure composition remain poorly understood (Li et al., 2016). A link between ultrastructure disruption and reduced expression of calcification proteins under elevated pCO$_2$ has been proposed (e.g., Huning et al., 2013; Fitzet et al., 2014b), although shell matrix protein candidate genes and the expression of carbonic anhydrase – the latter of which catalyzes the interconversion of bicarbonate and CO$_2$ – appear less affected by elevated CO$_2$ (Huning et al., 2013). Furthermore, our understanding of the interaction between protein expression, functioning, and CaCO$_3$ formation remains limited (see Glazier et al., 2020 for a recent study in corals). A number of models have been proposed including: changes in epitaxial growth, the formation of “mineral bridges” between crystals, or alteration of the c-axis (0001 plane) orientation during the earlier stages of crystal formation (Choi and Kim, 2000; Checa et al., 2006; Fryda et al., 2010; Olson et al., 2013; Maier et al., 2014). Clarification of changes in the physiological mechanisms including gene functioning and regulation under future climate scenarios may
shed light on the capacity of marine species, and especially calcifiers, to mitigate the effects of environmental change. 

Our mesocosm system uses fixed input concentrations of CO$_2$ and we do not attempt to modify these concentrations using feedback loops to control (fix) pH. Consequently, the ocean chemistry in our test system behaves in the same way as in situ seawater would i.e., a change in atmospheric pressure can lead to changes in surface water pCO$_2$ over short timescales (Burgers et al., 2017) and fluctuations in pH (see Lemasson et al., 2018). Fluctuations can also arise from close proximity to CO$_2$ vents (Hall-Spencer et al., 2008), the introduction of acidified water from coastal upwelling (Feely et al., 2008; Booth et al., 2012), and freshwater inputs (Aguilera et al., 2013; Waldbusser and Salisbury, 2014). Photosynthesis, respiration and the calcification and dissolution of biomineralized structures can also remove and release carbonate (CO$_3^{2-}$) ions (Anthony et al., 2011). In marine systems, individuals may therefore be exposed to rapid and acute perturbations of pH rather than long-term, chronic exposure to reduced pH as simulated in some mesocosm experiments with conditions varying over spatio-temporal scales. The impact of OA and warming may therefore vary greatly across spatial and temporal scales (e.g., Kurihara et al., 2007; Kurihara, 2008; Thomsen and Melzner, 2016; Fitzler et al., 2015a; Lemasson et al., 2018). Any response may also be context specific and dependent upon individual factors such as the acclimatory capacity and/or evolutionary origins of the target organism (Matoo et al., 2013 and references therein; Ramajo et al., 2019; Britton et al., 2019), capacity to perform biomineralization under variable environmental conditions (Day et al., 2000), or susceptibility to dissolution depending on the mineralogy of their shells (Melzner et al., 2011; Olson et al., 2013). Indeed, current evidence indicates that natural pH fluctuations can influence the impacts of OA on bio-mineralizing organisms (Wahl et al., 2016) or have no effect, even among species within the same genera (Frieder et al., 2014; Lemasson et al., 2017b, 2018).

Should change to the ultrastructural integrity and strength of individuals occur, there are potentially important ramifications for the fitness and survivorship of bivalve mollusks in the future. Given the common role of bivalves as ecosystem engineers (sensu Jones et al., 1994) in supporting biodiversity, a reduction in shell strength may lead to greater susceptibility to predation and wider multitrophic impacts. For instance, predators of Mytilus sp. (e.g., Carcinus maenas, Nucella lapillus) typically use crushing or boring to penetrate the shell (Elner, 1978; Sadler et al., 2018). Furthermore, predators of bivalve mollusks may themselves become susceptible to OA through impacts on fitness (Whiteley, 2011; Landes and Zimmer, 2012; Sadler et al., 2018), feeding structures (Landes and Zimmer, 2012) and predatory behavior (Dixson et al., 2010; Queiros et al., 2015; Sadler et al., 2018).

In recent years, there has been a burgeoning of studies assessing the effects of future climate change scenarios on the fitness, physiology, and provision of ecosystem services in a plethora of marine species at both the macroscale, and increasingly, microscale. Many earlier studies have shown detrimental impacts of OA and warming scenarios on species performance, but increasingly, others do not (see Clark et al., 2020). Here, our results indicate that future predictions of sea surface temperature are likely to have detrimental effects on the ultrastructure of Mytilus sp., but when combined with atmospheric pCO$_2$ concentrations predicted for 2100, those negative effects will be in part, ameliorated. These results further muddy the waters in terms of forming generic conclusions about the impact of climate change on marine organisms. It is important to recognize that in many instances, daily fluctuations in pCO$_2$ will be experienced by organisms in surface waters, that will change over spatial and temporal scales depending on environment. This presents challenges when making predictions of a general response of organisms to future climate conditions. We therefore propose that understanding how organisms respond to small and large-scale spatio-temporal fluctuations in future environmental conditions is the next challenge to predicting impacts and further disentangling the consequences of climate change.

**DATA AVAILABILITY STATEMENT**

The datasets presented in this article are not readily available because the dataset contains yet to be processed data and is therefore not freely available for use. Requests to access the datasets should be directed to AK, aknights@plymouth.ac.uk.

**AUTHOR CONTRIBUTIONS**

AK, NS, and AL devised the research and wrote the draft manuscript. MN and AL ran OAW experiments. MN and NS performed SEM analysis. All authors contributed to the article and approved the submitted version.

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**SUPPLEMENTARY MATERIAL**

The Supplementary Material for this article can be found online at: https://www.frontiersin.org/articles/10.3389/fmars.2020.567228/full#supplementary-material
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Conflict of Interest: The authors declare that the research was conducted in the absence of any commercial or financial relationships that could be construed as a potential conflict of interest.

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