Comparison of the bleaching susceptibility of coral species by using minimal samples of live corals

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ABSTRACT

In massive bleaching events (losing symbiotic algae from corals), more sensitive corals are bleached earlier than other corals. To perform a comparison of bleaching susceptibility within and across coral species, a simple quantitative method is required. Accordingly, we present a laboratory-based method for comparing the bleaching susceptibility of various coral species by using a standardized image analysis protocol. Coral fragments were sampled from the colonies of five species selected from Kenting, southern Taiwan, and maintained in the same aquarium tank with circulating seawater; 2 seawater temperature regimes were used (i.e., fast-heating program (FHP), with a heating rate of 1 °C per day; and slow-heating program (SHP), with a heating rate of 1 °C per 3 days). Each coral fragment was photographed periodically, and the colored images were subsequently converted to grayscale images and then digitally analyzed to determine the standardized grayscale values (G₀) by comparing with that of standard color strip. The G₀ of a sample at each time of photographing during bleaching was divided by the difference of G₀ between the acclimating and the same but completely bleached fragment to derive the relative grayscale (RG%) at a particular stage of bleaching; this is done for each coral fragment of a colony. The smaller the RG% of a coral fragment the closer it is approaching completely bleached condition. The level of decrease in RG% within a time series of images in each heating regime was used to establish a bleaching time index (BTI). The lower the BTI, the sooner to reach a defined bleaching level (e.g., 30%), this indicates the coral is more sensitive to thermal bleaching. In the experiment, we compared the bleaching susceptibility of the five species. Based on the proposed BTI, the five species were ranked in terms of bleaching susceptibility, and the rankings were identical between the two temperature regimes; three species in Pocilloporidae had lower BTI, whereas the hydrocoral Millepora species had the highest BTI. Within each heating regime, the BTI of different species were ranked and used to indicate susceptibility. In the FHP, the three Pocilloporidae species could be divided into two groups in terms of bleaching susceptibility. FHP not only displayed a higher differentiating capability on coal bleaching susceptibility than SHP, but also had a faster completion time, thus reducing the likelihood of unforeseen complications during the tank experiments. Our color-based method is easier and less effort-intensive than methods involving the assessment of zooxanthellae densities. Moreover, it requires much fewer replicates and all samples in one large tank (e.g., 300 L) for the studies considering multiple species comparisons. This method opens opportunities for studying the
effects of species types, acclimatization (e.g., seasons), and environmental factors other than temperature on coral bleaching.

Subjects Conservation Biology, Marine Biology
Keywords Coral, Bleaching susceptibility, Photoimaging, Image analysis

INTRODUCTION

Coral bleaching is primarily caused by the breakdown of coral-algal symbiosis under heat stress; the expulsion of symbiotic algae from thin coral tissues reveals the white carbonate skeleton underneath (Fitt et al., 2001; Douglas, 2003; Lesser & Farrell, 2004). In the context of global warming, the substantial spatial scale of coral bleaching with high mortality rates has caused a worldwide decline in coral coverage that has had a destructive impact on reef ecology (Hoegh-Guldberg et al., 2007; Jones et al., 2008). The increasing incidence of beyond-threshold seawater temperatures has increased the frequency of coral bleaching events such as that at the Great Barrier Reef in 2016 (Hughes et al., 2018). The large scale of coral bleaching highlights the catastrophic impact of global warming; this has thus stimulated considerable attention on the preservation of coral reefs.

Although the mechanisms underpinning coral bleaching are still not fully understood (reviewed in van Oppen & Lough, 2018), numerous efforts have been devoted to explore the possibility of enhancing the capacity of corals to survive at higher (future) temperatures. Strategies proposed for improving coral resistance to thermal stress entail applying temperature acclimatization and adaptation measures (e.g., Bellantuno, Hoegh-Guldberg & Rodriguez-Lanetty, 2012; Putnam & Gates, 2015; Majerova et al., 2021), searching for heat-tolerant genotypes in natural or artificial high-temperature environments (e.g., Coles & Riegl, 2012; Kao et al., 2018), applying probiotic microbe inoculation (Rosado et al., 2019), and using CRISPR/Cas9-editing techniques to mediate heat tolerance capability (Cleves et al., 2018). These strategies require examining the bleaching performance of corals in tank experiments under a temperature-increasing process to verify the thermal tolerability of the treated corals.

The rate of decrease in algal density in coral tissues is conventionally used as a measure to directly demonstrate the bleaching level of corals under thermal stress. Algal cell density, occasionally combined with chlorophyll a content from corals (e.g., Hoegh-Guldberg & Smith, 1989) or used along with a detailed measurement of photosynthetic processes and related physiological parameters (e.g., Jones et al., 1998; Fitt et al., 2001) enables monitoring the physiological condition of individual coral colonies under stress. However, deriving this measure necessitates specialized equipment and considerable effort, thus restricting the application of this approach. In addition to cell-counting methods, observer-based methods have been developed for field surveys of coral health through the use of color reference cards with multiple hues (Siebeck et al., 2006) or use of simple color reference strips supplemented with photographic images (Winters et al., 2009; Chow et al., 2016; Bryant et al., 2017). The feasibility and reliability of these methods have been confirmed using various approaches (e.g., Montano et al., 2010; Amid et al., 2018).
Observer-based methods that rely on a subjective color-based assessment of corals are quicker and less expensive than cell-counting methods but are potentially biased by the personal judgements of assessors. Such bias, however, could be addressed by applying images captured under a set of fixed conditions and then analyzing the images through computer software. Photography-based methods have considerably improved the operational efficiency of field surveys of coral bleaching events. Moreover, studies have demonstrated that image data were highly correlated with measurements of the algal cell density and chlorophyll content of corals, confirming the appropriateness of using image data to research bleaching conditions (Winters et al., 2009; Chow et al., 2016; Bryant et al., 2017). However, studies have yet to apply photography-based methods in tank experiments. Instead, they have generally applied cell-counting methods to compare the bleaching susceptibility of corals with different acclimatization histories or experimentally treated coral samples; such methods require a considerable quantity of living corals and more sophisticated handling efforts and procedures.

To fill this gap, the present study applied a photography-based method to distinguish coral bleaching under different conditions, such as different species and heating regimes, in a tank. We developed a bleaching time index (BTI), calculated from the changes in relative monotone grayscale of images with time, to compare coral bleaching susceptibility.

**MATERIALS AND METHODS**

A total of five species of corals from the shallow waters of Kenting, southern Taiwan (120°41′43″E, 21°59′5″N), were collected by scuba divers. The collection permit for this investigation was issued by the Kenting National Park Authority of Taiwan to KS. Four scleractinian corals (*Seriatopora caliendrum*, *Pocillopora verrucosa*, *Pocillopora damicornis*, and *Favites complanata*) and one zooxanthellate hydrocoral (*Millepora intricata*) were included in this study. For each coral species, five colony replicates were selected, and two coral fragments (4–8 cm per fragment) from each colony were harvested for the bleaching experiment. The species were transported to the laboratory and separated into two identical aquarium tanks, with each containing 300 L of seawater. Thus, each species contained five colony replications for each of two treatments, and the coral fragments of all five species were maintained in the same tank, for each temperature treatment.

Figure 1 illustrates the procedures of the study protocol. Specifically, the coral fragments harvested from all colonies of the five coral species were acclimatized at 27 °C in one of the tanks for 1 week prior to the experiment. The seawater in the aquarium tank was first circulated horizontally by two underwater pumps—each of which provided a flow rate of 5,200 L h⁻¹—and then filtered through several layers of aquarium filter pads using another pump (flow rate: 2,000 L h⁻¹). The corals were illuminated with photosynthetically active radiation (25–30 μmol m⁻² s⁻¹) under a 12:12-h light–dark regime. Heating of seawater was conducted with two heaters (500 w for each) controlled by a digital thermostat (ISTA tsb-958), which were located right in front of an underwater pump to quickly mix water and away from tested corals. During the heat treatment, one of the tanks was used to implement a fast-heating program (FHP; temperature increment...
rate = 1 °C per day), and the other tank was used to implement a slow-heating program (SHP; temperature increment rate = 1 °C per 3 days). The temperature setting was changed at noon each day, and approximately 1 h of heating was generally required to reach the target temperature. The water temperature was raised to 35 °C and maintained at that level until the entire coral fragments were pale.

During the heat treatment, one image of all colony replicates of each of the 5 species was captured twice a day (at 6:00 AM and 6:00 PM) under a constant photographing condition. All images were captured by an Olympus Tough TG-5 camera (Olympus, Tokyo, Japan) with an attached LED ring light (Weefine, Guangdong, China), aperture speed of 1/50 s, ISO level of 100, and aperture size of f/2.8; the camera was operated in macro mode, with the built-in camera flash turned off. The image acquisition process was completed when the coral fragments turned completely pale (i.e., no grayscale decreasing in 6 consecutive data points). All images were stored in JPG format for further bleaching level calculation.
The bleaching level of the coral fragments during the heat treatment was determined by deriving the changes in the relative grayscale values of a fixed square area in the images captured at every time point, as illustrated in Fig. S1. Adobe Photoshop CC (2017) was used to measure the grayscale value in each target square area of coral sample (G_C) and standard color strip (G_B for black and G_W for white). Before measuring the grayscale of a selected area, the color image was firstly converted to grayscale, and then, that of the selected area was averaged with the Photoshop software. To correct for the variations in conditions between image acquisition sessions, the G_C value derived for each image captured at each time point (6:00 AM and 6:00 PM) was first standardized with G_B and G_W to derive the standardized grayscale (G_0), as presented in Eq. (1). A step-by-step procedure for estimating G_0 was provided in Fig. S2. Subsequently, to measure color loss at each time point, the remaining color intensity, derived as G_0 - G_0^{bleach}, was normalized with the total grayscale change (G_0^{normal} - G_0^{bleach}) to derive the relative grayscale (RG%) of a sample at each time point during bleaching, as presented in Eq. (2). G_0^{bleach} represents the average G_0 of bleached coral fragments calculated from the images captured at the last 5 time points of heating experiment when the grayscale of coral fragment displays not further decrease; and G_0^{normal} was derived from the mean of the first 5 G_0 data points at 27 °C incubation. All preceding calculations were conducted for each fragment of each species; that is, the calculations of relative changes were affected only by the performance of the coral fragment.

\[
G_0 = \left( \frac{G_C - G_W}{G_B - G_W} \right)
\]  

(1)

\[
RG\% = \frac{G_0 - G_0^{bleach}}{G_0^{normal} - G_0^{bleach}} \times 100\%
\]  

(2)

The calculated RG% values were plotted against incubation time and analyzed through curve fitting in SigmaPlot 14.1 software. The best-fitting regression analysis indicated a modified Gaussian regression model revealing the highest regression coefficient (R^2). Parameters and R^2 values derived using the modified Gaussian equation for all 50 data sets are presented in Table S1; the R^2 values ranged from 0.9326 to 0.9978 (mean ± SD = 0.9856 ± 0.0125). The modified Gaussian regression equation was further used to estimate the time (in days) required for the grayscale to decrease during bleaching, and then applied to develop the BTI for evaluation of the bleaching rate and heat stress susceptibility of the various coral species. With the regression equation, SigmaPlot provided the corresponding X (incubation time point) and Y (RG%) values within the range of input data, which were used to calculate the incubation time (in days) at which RG% decreased by 10%, 20%, 30%, 40%, or 50% through interpolation. Because 10%, 20%, 30%, 40%, and 50% decreases in grayscale (color) values produced similar results in terms of the differentiation of bleaching susceptibility among the species, only results obtained at 10%, 30%, and 50% are presented in this study. We defined the time required to reduce the RG% value of coral samples during the heat treatment as the BTI. BTI values were compared between the coral species by using one-way analysis of variance followed by
Tukey’s test for post-hoc analyses. The appropriateness of using BTI to evaluate bleaching susceptibility was also examined with an analysis by Pearson correlation coefficient between the treatments of FHP and SHP.

RESULTS
When seawater was gradually heated to and maintained at 35 °C in both the FHP (1 °C per day) and SHP (1 °C per 3 days), the bleaching responses of the coral fragments harvested from the colonies of the 5 coral species varied (Fig. 2B–2F, 3B–3F, and 4. In Figs. 2

Figure 2 Bleaching responses of corals in the fast-heating program (FHP, 1 °C per day). (A) Temperature increment program, in which the seawater temperature was increased at a rate of 1 °C per day from 27 °C to 35 °C and maintained at 35 °C until the end of the experiment. (B) Seriatopora caliendrum, (C) Pocillopora verrucosa, (D) Pocillopora damicornis, (E) Favites complanata, and (F) Millepora intricata treated in one 300-L aquarium tank. Five color symbols represent coral fragments from different colonies.
and 3, we considered a 5% decrease in relative grayscale as the criterion to calculate the BTI in order to determine the time at which initial bleaching occurred. Accordingly, we observed that in the FHP, coral bleaching started on day 3.2 ± 0.3 for *S. caliendrum*, day 5.5 ± 0.4 for *P. verrucosa*, day 6.4 ± 1.1 for *P. damicornis*, day 6.8 ± 0.9 for *F. complanata*, and day 8.4 ± 0.8 for *M. intricata*. In the SHP, the bleaching started on day 7.9 ± 1.6 for *S. caliendrum*, day 9.9 ± 4.2 for *P. verrucosa*, day 7.9 ± 3.2 for *P. damicornis*, day 12.2 ± 2.6 for *F. complanata*, and day 22.5 ± 2.2 for *M. intricata*. Figures 2B–2F and 3B–3F also showed higher within-species variations in the days of starting bleaching as treated by SHP.
(CV%: 10.0–40.5%) than by FHP (CV%: 6.8–17.2%) in the five testing coral species. Notably, the susceptibility rankings of the coral replicates were the same between the FHP and SHP.

In Figs. 2 and 3, the hydrocoral *M. intricata* also exhibited the highest tolerance to temperature increases compared with the other four scleractinian coral species; specifically, it exhibited the longest endurance before the initiation of bleaching, tolerated the highest temperature before the initiation of bleaching, and exhibited the slowest bleaching rate. In addition, *M. intricata* displayed on average >40% increase in the relative grayscale before the coral started bleaching in the SHP, compared with ≤10% in the FHP and in both FHP and SHP for the other scleractinian coral species (Table S2).

Further to compare the BTI at different levels, as shown in Fig. 4, the five tested coral species displayed significantly varied response to the two treatments (*p* < 0.001, see ANOVA results in Table S3). However, the rankings of bleaching resistance among species determined by BTI10%, BTI30%, and BTI50% were identical in both heat treatment programs. A comparison of the 95% confidence intervals for the FHP and SHP (Fig. 4) revealed that eight out of 10 possible species pairs could be significantly distinguished in terms of bleaching resistance in the FHP; however, only 6 of 10 pairs could be significantly separated in terms of bleaching resistance in the SHP. A higher BTI indicates a greater bleaching resistance. Hence, in the FHP, the 5 coral species could be ranked (in descending order) into 4 groups as follows according to their bleaching resistance: *M. intricata* > *F. complanata* ≥ (*P. damicornis* = *P. verrucosa*) > *S. caliendrum* (Fig. 4 (ACE)). However, in the SHP, the species could be ranked (in descending order) into only 3 groups as follows according to their bleaching resistance: *M. intricata* > *F. complanata* > (*P. damicornis* = *P. verrucosa* = *S. caliendrum*) (Fig. 4 (BDF)). Furthermore, the bleaching performance of the five coral species treated in the FHP were also significantly correlated with those of the species treated in the SHP (Fig. 5). The Pearson correlation coefficients (*r*) for the BTI data between the SHP and FHP were 0.754 for BTI10% (*p* < 0.01), 0.895 for BTI30% (*p* < 0.01), and 0.898 for BTI50% (*p* < 0.01).

The variation of BTI method was further examined with the coefficients of variation (CV%) of BTI values derived from colony replicates of each species. As indicated in Table S4, the CV% levels of the BTI values at the 10%, 20%, 30%, 40%, and 50% cutoff levels were low in all the tested coral species treated by the FHP (mean ± SD = 8.3% ± 3.7%; *n* = 25) and two less sensitive species, namely *F. complanata* and *M. intricata*, by SHP (9.3% ± 4.0%; *n* = 10), but high in three thermal-sensitive species, namely *S. caliendrum*, *P. verrucosa*, and *P. damicornis*, by the SHP (24.7% ± 10.1%; *n* = 15). In summary, the results indicated that SHP would result in higher within-species variation in the BTI method when applied to heat sensitive coral species.

**DISCUSSION**

This study proposes a color-based protocol that entails the use of the BTI to evaluate coral bleaching susceptibility in tanks. The BTI is calculated from the changes in the relative grayscale values of a selected area in images of corals. In this study, the BTI10%, BTI30%, and BTI50% values (Figs. 4A, 4C and 4E) obtained in the FHP could be used to rank the 5
coral species into 4 groups of bleaching susceptibility. In contrast to the rankings observed in the FHP, the species treated in the SHP could be ranked into only three groups according to their bleaching susceptibility (Figs. 4B, 4D and 4F). Though the rankings were consistent between the SHP and FHP, slow heating rate did not increase resolution in differentiating bleaching susceptibility between the coral species. Consistent ranking
between the SHP and FHP demonstrated that the proposed BTI can be a fair tool used to assess natural processes. Besides that, monitoring bleaching process by the SHP displayed two features different from that by FHP. First, SHP resulted in higher within-species variations in estimating BTI values of 3 heat sensitive species, *S. caliendrum*, *P. verrucosa*, and *P. damicornis* than FHP did (Table S4). Within species, different susceptibility levels may be caused by both genetic and/or local environmental factors, which might be only observed by SHP. However, to the heat sensitive species, the likelihood of unforeseen complications occurring in tank experiment also can’t be ruled out due to longer incubation time in tank condition. Second, to heat resistant species, like hydrocoral *M. intricata* in this study, there was a higher percentage of increase in relative grayscale in the SHP (40% in average) than in the FHP (10% in average) before it declined relative grayscale below 100% (Figs. 2F and 3F, Table S2). The increases in relative grayscale in the initial period of bleaching trials might be due to the accumulation of expelled symbiotic algae in the topical area of coral tissue. The high initial rise of RG% of *M. intricata* might be partly attributable to their heat resistant nature resulting in a slower mode of symbiotic algae releasing, and partly to their porous skeleton which made the releasing of symbiotic algae more slowly.

Based on our results, the FHP are superior to the SHP in comparing the bleaching susceptibility coral species. Nevertheless, it remains to be tested whether the BTI could also produce the same rankings for species subjected to other methods of bleaching induction. The possible effects of zooxanthellae, feeding, and light can also be tested under laboratory conditions using this method.

The bleaching susceptibility levels of the scleractinian coral species observed in this study using the proposed BTI are comparable to those reported by previous studies (e.g., Loya et al., 2001; Keshavmurthy et al., 2014); however, the results observed for the branching hydrocoral coral *M. intricata* are not comparable to those in the literature.
Previous studies conducting field surveys have indicated that *Millepora* species are thermally sensitive (*Loya et al., 2001; Dias & Gondim, 2016; Teixeira et al., 2019; Duarte et al., 2020*), but we discovered that *M. intricata* was the most thermally resistant species in this study. This conflict might be derived from two possibilities, method bias and/or the other biological variations. The essential difference between BTI method and field survey on *M. intricata* is that the former provided time-series quantitative observations of the same coral fragments and that the bleaching was relative to the acclimation periods. On the contrary, the evidence of heat sensitive in *Millepora* coral was all based on visual observation during field surveys. Field survey data have been limited to snapshots of multiple species taken at the same time; the assessments of bleaching were qualitative, although many colonies were selected for sampling. Of course, we cannot exclude other biological factors that might affect coral bleaching susceptibility, such as thermal sensitivity of algal symbionts (*Sampayo et al., 2008; Stat & Gates, 2011; Howells et al., 2012; Hsu et al., 2012; Silverstein, Correa & Baker, 2012; Keshavmurthy et al., 2014*), tissue thickness and stress enzyme activities of the coral host (*Loya et al., 2001; Fitt et al., 2009; Wang et al., 2019*), with or without probiotic microbes association (*Rosado et al., 2019* and the references therein), and life histories with or without acclimatization or adaptation to higher temperature (*Howells et al., 2016; DeCarlo et al., 2019; Wang et al., 2019; Barott et al., 2021*). More studies are required to clarify the effects of these factors, and we suggest the methods developed here is more effective and efficient.

Imaging methods have several advantages over commonly used cell-counting methods (Table 1). First, imaging methods do not require sacrificing coral samples for cell counting; therefore, increasing sampling intervals does not require proportionally increasing the number of coral fragment samples, as is the case in cell-counting methods.

| Requirements and advantages | Methodology |
|-----------------------------|-------------|
|                             | Counting algal cells (CZ) | Photography image (PI) |
| Sampling area               | Hard to obtain accurate area, especially for branching corals | No need to calculate sampling area |
| Number of nubbins per colony| Number of nubbins per colony needs to increases as number of observations increase from the beginning of the experiment to the end of bleaching | Same coral nubbins for multiple times of non-destructive photography. Sample numbers could be easily increased without collecting more/larger coral fragments |
| Equipment                   | Water pick, centrifuge, microscope, hemocytometer etc. | Camera, light, color card, software, e.g., Photoshop etc. |
| Skills                      | Cell-biology training | Digital photography training |
| Lab facility                | Proportionally larger space according to: species numbers, and numbers of observation per spp. | Proportionally larger space according to species numbers, only |
| Raw data                    | Y: Cell densities | Y: % in grayscale |
| Sampling assumptions        | Homogeneity among nubbins within a colony in cell numbers | No such assumption is needed since the same nubbins were observed continually |
| Numbers of data point       | Constrained by the numbers of nubbins available for assay | Flexible; depending on intervals between shots |
| Numbers of species compared simultaneously | Less | More, we did 5 species comparison in a 300 L aquarium tank |

Table 1 Comparisons of imaging and algal cell-counting methods for quantifying coral bleaching among species.
For example, if our experiment had been conducted using a cell-counting method, 60 times of replicate of coral fragments (i.e., 300) would have been required for each species to generate the same number of observations.

Second, due to much smaller sample sizes required for a species, many species could be accommodated in one tank. This effectively avoided the tank effect, which we are not really interested any way, and also reduced the potential complication of pseudo-replication in experimental design (Cornwall & Hurd, 2016).

Third, imaging methods have lower technological requirements and are less labor intensive than cell-counting methods. Unlike an underwater imaging method used a decade ago (Winters et al., 2009), current imaging approaches are easier with the increased availability of automatically controlled underwater cameras, such as that used in this study. Furthermore, compared with labor-intensive cell-counting methods, imaging methods provide additional data points at low marginal costs.

Fourth, imaging methods provide a lower level of data bias than do cell-counting methods when used for intertaxa comparisons of bleaching performance under the same controlled conditions. Cell-counting methods usually underestimate algal densities in coral species with porous skeletons, such as Millepora spp.; in such species, complete recovery of algal cells from coral specimens by using methods commonly applied to scleractinian corals (e.g., water-pik or air brush; Edmunds, 1999) is almost impractical. By contrast, the imaging method used in this study derived only monotone (or grayscale) color changes during the bleaching process, which can provide the same observation of this process as the naked eyes. Although the colors of non coral backgrounds (e.g., that of endolithic algae or the supporting materials for coral fixation) may distort the actual grayscale levels of coral samples in an image, such backgrounds can be removed from the images during the calculation of relative grayscale values within a sample. Therefore, our protocol affords more freedom in the selection of a region of interest in an image for calculating grayscale values, which reduces the number of coral fragment replication points to one. This is an advantage over the protocol proposed by Chow et al. (2016), in which more than 100 coral fragment areas are required for calculating grayscale values.

Imaging methods, nevertheless, have disadvantages in the monitoring of coral bleaching because they require a series of images, including images of the normal and the bleached stages of corals. Consequently, imaging methods are not suitable for directly assessing bleaching conditions in nature. Also, converting JPG image into grayscale is a nonlinear compression process, which might affect greyscale quantification. However, according to our results, this defect might only have minor impact on the final comparisons.

CONCLUSIONS

In summary, we developed a BTI for comparing the bleaching susceptibility of corals based on imaging method. The method used coral fragments efficiently and the small numbers of fragments themselves allow multiple species comparisons in the same tanks. This characteristic enable investigations of a wide variety of factors causing coral bleaching in the laboratory.
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Competing Interests
The authors declare that they have no competing interests.

Author Contributions
• Jih-Terng Wang conceived and designed the experiments, performed the experiments, analyzed the data, prepared figures and/or tables, authored or reviewed drafts of the paper, and approved the final draft.
• Chi-Wei Chu performed the experiments, analyzed the data, prepared figures and/or tables, and approved the final draft.
• Keryea Soong conceived and designed the experiments, analyzed the data, authored or reviewed drafts of the paper, and approved the final draft.

Field Study Permissions
The following information was supplied relating to field study approvals (i.e., approving body and any reference numbers):
Collection of coral was approved by Kending National Park.

Data Availability
The following information was supplied regarding data availability:
The raw measurements are available in the Supplemental Files.

Supplemental Information
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REFERENCES

Amid C, Olstedt M, Gunnarsson JS, LeLan H, TranThiMinh H, VandenBrink PJ, Hellström M, Tedengren M. 2018. Additive effects of the herbicide glyphosate and elevated temperature on the branched coral Acropora formosa in Nha Trang, Vietnam. Environmental Science and Pollution Research International 25:13360–13372 DOI 10.1007/s11356-016-8320-7.

Barott KL, Huffmyer AS, Davidson JM, Lenz EA, Matsuda SB, Hancock JR, Innis T, Drury C, Putnam HM, Gates RD. 2021. Coral bleaching response is unaltered following acclimatization to reefs with distinct environmental conditions. Proceedings of the National Academy of Sciences of the United States of America 118:e202535118 DOI 10.1073/pnas.202535118.

Bellantuno AJ, Hoegh-Guldberg O, Rodriguez-Lanetty M. 2012. Resistance to thermal stress in corals without changes in symbiont composition. Proceedings of the Royal Society B: Biological Sciences 279:1100–1107 DOI 10.1098/rspb.2011.1780.

Bryant DEP, Rodriguez-Ramirez A, Phinn S, González-Rivero M, Brown KT, Neal BP, Hoegh-Guldberg O, Dove S. 2017. Comparison of two photographic methodologies for collecting and analyzing the condition of coral reef ecosystems. Ecosphere 8(10):e01971 DOI 10.1002/ecs2.1971.

Chow MH, Tsang RHL, Lam EKY, Ang P Jr. 2016. Quantifying the degree of coral bleaching using digital photographic technique. Journal of Experimental Marine Biology and Ecology 479:60–68 DOI 10.1016/j.jembe.2016.03.003.

Cleves PA, Strader ME, Bay LK, Pringle JR, Matz MV. 2018. CRISPR/Cas9-mediated genome editing in a reef-building coral. Proceedings of the National Academy of Sciences of the United States of America 115(20):5235–5240 DOI 10.1073/pnas.1722151115.

Coles SL, Riegl BM. 2012. Thermal tolerances of reef corals in the Gulf: a review of the potential for increasing coral survival and adaptation to climate change through assisted translocation. Marine Pollution Bulletin 72(2):323–332 DOI 10.1016/j.marpolbul.2012.09.006.

Cornwall CE, Hurd CL. 2016. Experimental design in ocean acidification research: problems and solutions. ICES Journal of Marine Science 73(3):572–581 DOI 10.1093/icesjms/fsv118.

Dias TLP, Gondim AI. 2016. Bleaching in scleractinians, hydrocorals, and octocorals during thermal stress in a northeastern Brazilian reef. Marine Biodiversity 46(1):303–307 DOI 10.1007/s12526-015-0342-8.

DeCarlo TM, Harrison HB, Gajdzik L, Alaguarda D, Rodolfo-Metalpa R, D’Olivo J, Liu G, Patalwala D, McCulloch MT. 2019. Acclimatization of massive reef-building corals to consecutive heatwaves. Proceedings of the Royal Society B: Biological Sciences 286(1898):20190235 DOI 10.1098/rspb.2019.0235.

Douglas AE. 2003. Coral bleaching—how and why? Marine Pollution Bulletin 46:385–392 DOI 10.1016/S0025-326X(03)00037-7.

Duarte GAS, Villela HDM, Deocleciano M, Silva D, Barro A, Cardoso PM, Villela CLS, Rosado P, Messias CSMA, Chacon MA, Santoro EP, Olmedo DB, Szpilman M, Rocha LA, Sweet M, Peixoto RS. 2020. Heat waves are a major threat to turbid coral reefs in Brazil. Frontiers in Marine Science 7:179 DOI 10.3389/fmars.2020.00179.

Edmunds PJ. 1999. The role of colony morphology and substratum inclination in the success of Millepora alcicornis on shallow coral reefs. Coral Reefs 18:133–140 DOI 10.1007/s003380050167.

Fitt WK, Brown BE, Warner ME, Dunne RP. 2001. Coral bleaching: interpretation of thermal tolerance limits and thermal thresholds in tropical corals. Coral Reefs 20:51–65 DOI 10.1007/s0033800100146.

Fitt WK, Gates RD, Hoegh-Guldberg O, Bythell JC, Jatkar A, Grottoli AG, Gomez M, Fisher P, Lajenuesse TC, Pantos O, Iglesias-Prieto R, Franklin DJ, Rodrigues LJ, Torregiani JM,
van Woesik R, Lesser M. 2009. Response of two species of Indo Pacific corals, Porites cylindrica and Stylophora pistillata, to short-term thermal stress: the host does matter in determining the tolerance of corals to bleaching. *Journal of Experimental Marine Biology and Ecology* **373**:102–110 DOI 10.1016/j.jembe.2009.03.011.

Hsu C-M, Keshavmurthy S, Denis V, Kuo C-Y, Wang J-T, Meng P-J, Chen CA. 2012. Temporal and spatial variations in symbiont communities of catch bowl coral *Isopora palifera* (Scleractinia: Acroporidae) on Reefs in Kenting National Park, Taiwan. *Zoological Studies* **51**:1343–1353.

Hoegh-Guldberg O, Smith GJ. 1989. The effect of sudden changes in temperature, light and salinity on the population-density and export of zooxanthellae from the reef corals *Stylophora pistillata* Esper and *Seriatopora hystrix* Dana. *Journal of Experimental Marine Biology and Ecology* **129**:279–303 DOI 10.1016/0022-0981(89)90109-3.

Hoegh-Guldberg O, Mumby PJ, Hooten AJ, Steneck RS, Greenfield P, Gomez E, Harvell CD, Sale PF, Edwards AJ, Caldeira K, Knowlton N, Eakin CM, Iglesias-Prieto R, Muthiga N, Bradbury RH, Dubi A, Hatziolos ME. 2007. Coral reefs under rapid climate change and ocean acidification. *Science* **318**:1737–1742 DOI 10.1126/science.1152509.

Howells EJ, Beltran VH, Larsen NW, Bay IK, Willis BL, van Oppen MJH. 2012. Coral thermal tolerance shaped by local adaptation of photosymbionts. *Nature Climate Change* **2**:116–120 DOI 10.1038/nclimate1330.

Howells EJ, Abrego D, Meyer E, Kirk NL, Burt JA. 2016. Host adaptation and unexpected symbiont partners enable reef-building corals to tolerate extreme temperatures. *Global Change Biology* **22**:2702–2714 DOI 10.1111/gcb.13250.

Hughes TP, Kerry JT, Baird AH, Connolly SR, Dietzel A, Eakin CM, Heron SF, Hoey AS, Hoogenboom MO, Liu G, McWilliam MJ, Pears RJ, Pratchett MS, Skirving WJ, Stella JS, Torda G. 2018. Global warming transforms coral reef assemblages. *Nature* **556**:492–496 DOI 10.1038/s41586-018-0041-2.

Jones RJ, Hoegh-Guldberg O, Larkum AWD, Schreiber U. 1998. Temperature-induced bleaching of corals begins with impairment of the CO₂ fixation mechanism in zooxanthellae. *Plant Cell and Environment* **21**:1219–1230 DOI 10.1046/j.1365-3040.1998.00345.x.

Jones AM, Berkelmans R, van Oppen MJH, Mieog JC, Sinclair W. 2008. A community change in the algal endosymbionts of a scleractinian coral following a natural bleaching event: field evidence of acclimatization. *Proceedings of the Royal Society B: Biological Sciences* **275**:1359–1365 DOI 10.1098/rspb.2008.0069.

Kao KW, Keshavmurthy S, Tsao CH, Wang JT, Chen CA. 2018. Repeated and prolonged temperature anomalies negate Symbiodiniaceae genera shuffling in the coral *Platygyra verweyi* (Scleractinia; Merulinidae). *Zoological Studies* **57**:e55 DOI 10.6620/ZS.2018.57-55.

Keshavmurthy S, Meng P-J, Wang J-T, Kuo CY, Yang SY, Hsu CM, Gan CH, Dai CF, Chen CA. 2014. Can resistant coral-*Symbiodinium* associations enable coral communities to survive climate change? A study of a site exposed to long-term hot water input. *PeerJ* **2**:e327 DOI 10.7717/peerj.327.

Lesser MP, Farrell JH. 2004. Exposure to solar radiation increases damage to both host tissues and algal symbionts of corals during thermal stress. *Coral Reefs* **23**:367–377 DOI 10.1007/s00338-004-0392-z.

Loya Y, Sakai K, Yamazato K, Nakano Y, Sambali H, van Woesik R. 2001. Coral bleaching: the winners and the losers. *Ecology Letters* **4**:122–131 DOI 10.1046/j.1461-0248.2001.00203.x.
Majerova E, Carey FC, Drury C, Gates RD. 2021. Preconditioning improves bleaching tolerance in the reef-building coral *Pocillopora acuta* through modulations in the programmed cell death pathways. *Molecular Ecology* 30(14):3560–3574 DOI 10.1111/mec.15988.

Montano S, Seveso D, Galli P, Obura DO. 2010. Assessing coral bleaching and recovery with a colour reference card in Watamu Marine Park, Kenya. *Hydrobiologia* 655:99–108 DOI 10.1007/s10750-010-0407-4.

Putnam HM, Gates RD. 2015. Preconditioning in the reef-building coral *Pocillopora damicornis* and the potential for trans-generational acclimatization in coral larvae under future climate change conditions. *Journal of Experimental Biology* 218:2365–2372 DOI 10.1242/jeb.123018.

Rosado PM, Leitel DCA, Duarte GAS, Chaloub RM, Jospiin G, da Rocha UN, Saraiva JP, Dini-Andreote F, Eisen JA, Bourne DG, Peixoto RS. 2019. Marine probiotics: increasing coral resistance to bleaching through microbiome manipulation. *ISME Journal* 13(4):921–936 DOI 10.1038/s41396-018-0323-6.

Sampayo EM, Ridgeway T, Bongaerts P, Hoegh-Guldberg O. 2008. Bleaching susceptibility and mortality of corals are determined by fine-scale differences in symbiont type. *Proceedings of the National Academy of Sciences* 105(30):10444–10449 DOI 10.1073/pnas.0708049105.

Siebeck UE, Marshall NJ, Kluter A, Hoegh-Guldberg O. 2006. Monitoring coral bleaching using a colour reference card. *Coral Reefs* 25(3):453–460 DOI 10.1007/s00338-006-0123-8.

Silverstein RN, Correa AMS, Baker AC. 2012. Speciﬁcity is rarely absolute in coral-algal symbioses: implications for coral response to climate change. *Proceedings of the Royal Society B: Biological Sciences* 279(1738):2609–2618 DOI 10.1098/rspb.2012.0055.

Stat M, Gates RD. 2011. Clade D Symbiodinium in scleractinian corals: a “nugget” of hope, a selﬁsh opportunist, an ominous sign, or all of the above? *Journal of Marine Biology* 2011:730715 DOI 10.1155/2011/730715.

Teixeira CD, Leitão RLL, Ribeiro FV, Neves LM, Bastos AC, Pereira-Filho GH, Kampel M, Salomon PS, Sá JA, Falsarella LN, Amario M, Abieri ML, Pereira RC, Amado-Filho GM, Moura RL. 2019. Sustained mass coral bleaching (2016-2017) in Brazilian turbid-zone reefs: taxonomic, cross-shelf and habitat-related trends. *Coral Reefs* 38(4):801–813 DOI 10.1007/s00338-019-01789-6.

van Oppen MJH, Lough JM. 2018. Synthesis: coral bleaching: patterns, processes, causes and consequences. In: van Oppen M, Lough J, eds. *Coral Bleaching: Ecological Studies (Analysis and Synthesis).* Vol. 233. Cham: Springer.

Wang J-T, Wang Y-T, Keshavmurthy S, Meng P-J, Chen CA. 2019. The coral *Platygyra verweyi* exhibits local adaptation to long-term thermal stress through host-specific physiological and enzymatic response. *Scientific Reports* 9:13492 DOI 10.1038/s41598-019-49594-z.

Winters G, Holzman R, Blekhman A, Beer S, Loya Y. 2009. Photographic assessment of coral chlorophyll contents: implications for ecophysiological studies and coral monitoring. *Journal of Experimental Marine Biology and Ecology* 380:25–35 DOI 10.1016/j.jembe.2009.09.004.