First report of lionfish prey from Western Florida waters as identified by DNA barcoding

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ABSTRACT

DNA barcoding was used to identify prey fragments recovered from the stomachs of lionfish harvested during the 2016 Sarasota Lionfish Derby. A total of 305 prey fragments were recovered from 50 stomachs (mean = 4.6 per stomach), of which 184 (60.3%) fragments could be identified to either species or genus when Cytochrome Oxidase I (COI) sequences were queried against the Barcode of Life Database. We identified 21 fish prey species which represented fourteen families and accounted for 95.7% of genetically identifiable prey items. The remaining prey items identified corresponded to six crustacean species. The four most common prey taxa in lionfish stomachs were Ptereleotris calliura (24.3%), an unidentified Microgobius species (20.4%), Diplectum formosum (14.3%), and Apogon aurolineatus (12.2%). The most frequently observed crustacean species, Metapenaeopsis goodei, was found in only three stomachs (6.1%). We also report eleven taxa as putative novel lionfish prey species, most of which are common in Florida waters. Sixteen prey items were identified as lionfish (P. volitans); however, it was not definitive whether these detections were due to cross contamination or cannibalization. This represents the first report of lionfish diets from Florida waters in the Eastern Gulf of Mexico based on barcoding efforts. Our results are largely congruent with previous COI barcoding based studies of lionfish diets, indicating these predators to be generalists exhibiting preferences for specific prey traits but with regional differences in their diets.

INTRODUCTION

Native to the Indo-Pacific, lionfish (Pterois volitans and P. miles) have quickly become established in reefs and coastal habitats throughout the Western Atlantic since their introduction in the 1980s. Their settlement in this region has raised concerns of severe detrimental effects on local fauna, as lionfish have been shown to drastically reduce recruitment of native coral reef fishes (Albins & Hixon, 2008), exhibit high prey consumption rates (Green, Akins & Côté, 2011), prey on native fishes and invertebrates (Green, Akins & Côté, 2011; Côté et al., 2013; Dahl & Patterson, 2014; Harms-Tuohy, Schizas & Appeldoorn, 2016; Dahl et al., 2017) including the critically endangered wrasse...
Halichoeres socialis (Rocha et al., 2015), and drive declines in reef fish populations (Green et al., 2012; Albins, 2013, 2015). Controlled experiments exemplify these concerns, as they have shown lionfish to drive abundance declines and loss of species in Bahamian reefs at higher rates than native predators of similar sizes (Albins, 2013) and population declines of >50% in native species in experimental replicates (Ingeman, 2016). These findings underscore the importance of defining the prey preferences of lionfish in the Western Atlantic, specifically in areas where lionfish are relatively understudied.

In the Gulf of Mexico, settlement and spread of lionfish is thought to only have occurred over the past ten years. The first record of lionfish in the Gulf of Mexico (2006) is from a carcass found off the coast of Pinellas County, FL and is believed to represent a short-term individual (Schofield, 2010). In 2009 however, two lionfish were captured in the western Gulf of Mexico off the coast of Yucatan, Mexico (Aguilar-Perera & Tuz-Sulub, 2010), a record thought to be the first resulting from larval dispersal into the Gulf. Within a year of this capture, lionfish were reported along the U.S. coastline of the Gulf of Mexico. In August of 2010, lionfish were reported from waters near Cortez in Manatee County, FL (Schofield, 2010). By September of 2010, lionfish had been confirmed in the northern Gulf of Mexico, off the coasts of Florida and Alabama (Schofield, 2010).

Although widely studied in other areas of the Western Atlantic, particularly in the Caribbean, research into the predatory habits within the Gulf of Mexico remains comparatively sparse perhaps as a result of their relatively shorter invasive history. To date, documentation of lionfish prey species in the region have come from the Yucatan (Valdez-Moreno et al., 2012) and the northern Gulf of Mexico (Dahl & Patterson, 2014; Dahl et al., 2017, 2018). These studies, though varied in their sampling regime and identification approaches (i.e., molecular v. visual), indicate that lionfish in the region feed primarily on small fishes and to a lesser extent on some crustacean species with some variation associated with lionfish ontogeny. These findings are largely concordant with DNA barcoding results of lionfish prey items found in other Western Atlantic habitats, including Belize (Rocha et al., 2015), Puerto Rico (Harms-Tuohy, Schizas & Appeldoorn, 2016), and The Bahamas (Côté et al., 2013). Despite several identified taxa common among these studies, regional differences appear to exist between sites as each study has identified novel lionfish prey species. Thus, surveys of lionfish stomach contents in yet-to-be studied areas are needed to further our understanding of lionfish prey diets and preferences. This research would be valuable for understanding the local impacts of lionfish predation (Mizrahi et al., 2017) and for informing local management of lionfish (Bogdanoff et al., 2018). Towards this goal, we applied Cytochrome Oxidase I (hereafter COI) barcoding to prey fragments found in the stomachs of lionfish collected during the 2016 Sarasota Lionfish Derby where no previous studies on lionfish diets had been conducted and near the area where lionfish presence was first documented in the Gulf of Mexico.

**MATERIALS AND METHODS**

**Specimen collection**

Lionfish were collected during the 2016 Sarasota Lionfish tournament in artificial habitat constructed of limestone rocks piled over the Gulfstream natural gas pipeline.
approximately 30 miles west of St. Petersburg, FL in 30 m water depth (Fig. 1). The limestone rubble topping sections of the pipeline creates an artificial reef in an otherwise unconsolidated sandy substrate with occasional natural hardbottom features. Lionfish were stored on ice until the removal of stomachs which was completed within 24 h of capture. Upon removal, stomachs were placed in individually labeled Whirl-Pak bags and stored at −20 °C until dissections were to be performed. Prior to dissections, the Total Length (hereafter TL) of each lionfish was recorded. Collections were carried out in accordance with all applicable local, state, and federal laws and regulations.

**Dissections and molecular work**

Prey items from individual stomachs were isolated following the dissection protocol of Green, Akins & Morris (2012). During each dissection, all recognizable prey item fragments (e.g., spines, vertebrae, carapace fragments, whole fish, etc.) in each stomach were separated and inspected under a dissecting microscope. For each recognizable prey fragment, we estimated the rank of digestion (hereafter RD) per Green, Akins & Morris (2012), with prey fragments exhibiting no to minor degradation and which could be visually identified to species were assigned low RD scores (e.g., 1 and 2); highly digested items not possible to visually identify to genus or species were assigned higher scores (e.g., 4 and 5). Prey fragments were stored individually in 70% ethanol.

Total genomic DNA was extracted from each prey fragment using Zymo’s Quick-gDNA MiniPrep following standard protocol instructions. Smaller prey fragments (e.g., single...
spines) were used directly in DNA extractions. Fragments large enough to sub-sample (e.g., soft tissue), were dissected to obtain a small (2–5 mm) internal tissue sample as per Côté et al. (2013). Cross-contamination of samples was avoided by flame sterilizing tools after rinsing in 70% ethanol and as per Dahl et al. (2017). A 313-bp section of the COI gene was then amplified using the mlCOIintF forward (5′-GGWACWGGWTGAACWGT WTAYCCYCC) and jgHCO2198 reverse (5′-TAIAICYTGIGRTGICRAARYA) primers designed for the study of prey gut items by Leray et al. (2013) following the reaction conditions used by Harms-Tuohy, Schizas & Appeldoorn (2016). PCR mixtures consisted of 10 µL of QIAGEN Multiplex PCR mix, 7.8 µL of PCR-grade water, 0.6 µL of 10 µm of each primer, and 1 µL of genomic DNA with PCR reactions consisting of 16 initial cycles of denaturation for 10 s at 95 °C, annealing for 30 s at 62 °C, and extension for 60 s at 72 °C, followed by 25 cycles with identical denaturation and extension steps but annealing at 46 °C, and lastly a final extension cycle of 6 min at 72 °C. For a subset of samples, we amplified a larger (658-bp) section of the COI gene using the LCO-1490 forward (5′-GGTCAACAAATCATAAAGATATTGG) and HCO-2198 reverse (5′-TAAACTTCAGGGTGACCAAAAAATCA) primers designed by Folmer et al. (1994) using PCR mixtures were as described above, with PCR reactions consisting of 35 cycles of denaturation for one minute at 95 °C, annealing for one minute at 40 °C, and extension for one and a half minute at 72 °C, with a final extension cycle of seven minutes at 72 °C. Positive amplicons were identified using 1% agarose gels stained using SYBR Safe (Life Technologies Corporation) prior to sequencing at the University of Arizona Genetics Core.

Sequences were visually inspected, assembled, and edited (i.e., primers removed) using Geneious v8.0.5. Short sequences (final length <200-bp), those displaying signs of cross-amplification or evidence suggestive of pseudo-genes (e.g., early stop codons) were not included in our analyses.

The Barcode of Life Database (BOLD) (Ratnasingham & Hebert, 2007) was used to establish the identity of each successfully sequenced prey fragment. BOLD searches were carried out using the “All Barcode Records on BOLD” option during September of 2018 with taxonomic identifications based on the following criteria. Queries which produced matches of >98% similarity were considered confident species-level identifications. Those which resulted in similarity scores of 90–98% scores were considered confident genus-level identifications provided that: (a) the queried sequence clustered within a monophyletic clade composed solely of members of a single genus in the Tree Based identifications provided by BOLD, and (b) no other genera produced similar match scores. Similar conditions were used to identify sequences producing matches of 80–90% to families. These categories were chosen because they largely correspond with the average barcoding gaps reported for teleost fish in various regions (Zhang & Hanner, 2012; Knebelsberger et al., 2014; Chang et al., 2017). Sequences not confidently identified to species but resulting in a confident genus or family level identification were combined into groups of highly similar sequences (>98% similarity) using the assembly tool of Geneious R8 to ensure that no multiple unidentified species were present in our dataset. These groups are considered as operating taxonomic units for the rest of the study (OTUs). All taxonomic
identifications were checked against GenBank’s BLASTn results; however, we observed no discrepancies.

The median, and mode RD were calculated for all prey fragments, those successfully amplified and sequenced, and those which failed to amplify. We also estimated the abundance of each prey species as the number of lionfish stomachs which contained a given prey species and the frequency of occurrence of each prey species as the percentage of stomachs which contained a given prey species.

RESULTS
A total of 70 stomachs were dissected from lionfish ranging in size from 140 to 400 mm total length (TL; mean\(\text{TL} = 271.71\) mm (SD = 69.86); median = 270 mm; mode = 269 mm; Interquartile Range: 221.5–332.5 mm). Twenty of the stomachs did not contain any observable prey fragments (28.6%). These individuals did not statistically differ in size from those which contained prey fragments (\(t(68) = -1.48987; p\)-value = 0.1410). In the remaining fifty stomachs (71.4%), we recovered 305 prey fragments (per stomach mean = 6.1, median = 4, mode =1). Of these, 184 prey fragments from 49 lionfish stomachs produced sequences meeting the minimum quality criteria (e.g., short sequences, evidence of cross-amplification). The remaining 121 fragments either failed to amplify despite several PCR attempts or produced sequences of low quality. This represents a 60.3% success rate of amplification. Most fragments were highly digested (RD: median = 4.5; mode = 5), preventing effective visual identification. Fragments that failed to amplify and/or produced low quality sequences were slightly more digested (median RD = 4.5; mode = 5) than those successfully amplified (median RD = 4; mode = 4); however, this difference was not significant (\(t(303) = 0.3253, p = 0.7452\)).

BOLD-based identifications produced species or genus level-matches for all but three sequences; 67.4% of all sequences were identified to species and 31.5% to a genus. Of these, sixteen fragments were identified as lionfish (\(P. volitans\)). Although cannibalism is known to occur in lionfish in the Western Atlantic (Valdez-Moreno et al., 2012; Dahl et al., 2018), we were unable to determine whether sequences identified as \(P. volitans\) were the result of cannibalism or cross-contamination. We thus removed these sequences from further analyses. Remaining fragments represented 21 fish species and six crustacean species. Fish prey species accounted for most obtained sequences (95.7%) and were present in the majority of lionfish stomachs (\(N_{\text{stomachs}} = 41\)). The most commonly observed fish species were \(Ptereleotris calliura\) (Blue goby; \(N_{\text{stomachs}} = 12\); 24.3% of stomachs containing prey fragments), an unidentified species of \(Microgobius\) (\(N_{\text{stomachs}} = 10\); 20.4%), \(Diplectrum formosum\) (Sand perch; \(N_{\text{stomachs}} = 7\); 14.3%), \(Apogon aurolineatus\) (Bridle cardinalfish; \(N_{\text{stomachs}} = 6\); 12.2%), \(Heamulon aurolineatum\) (Tomtate grunt; \(N_{\text{stomachs}} = 5\); 10.2%). At the family level, the 21 fish species represented fourteen families. The most frequently observed fish families were Microdesmidae (Dartfishes; \(N_{\text{stomachs}} = 12\); 24.5%), Apogonidae (Cardinalfishes; \(N_{\text{stomachs}} = 11\); 22.4%), Gobiidae (True gobies; \(N_{\text{stomachs}} = 11\); 22.4%), and Serranidae (Seabasses and Groupers; \(N_{\text{stomachs}} = 8\); 16.3%). Crustacean species belonged to six crustacean families all in the Class Malacostraca. With the
exception of *Metapenaeopsis goodei* (Caribbean velvet shrimp; $N_{stomachs} = 3; 6.1\%$), all crustacean species were observed in a single stomach. Details of the occurrence of each species and family found in lionfish stomachs are presented in Table 1.

Table 1  List of prey species identified from lionfish (*Pterois* spp.) stomachs collected in waters off St. Petersburg, FL. Numbers outside parentheses indicate the number of stomachs in which each organism was found. Asterisks denote putative novel prey species. Bold terms indicate higher taxonomic levels.

| Phylum Chordata | Class Actinopterygii | Phylum Arthropoda | Class Malacostraca |
|-----------------|----------------------|------------------|-------------------|
| Family Apogonidae | 11 (22.4\%) | Family Alpheidae | 1 (2.0\%) |
| Apogon aurolineatus | 6 (12.2\%) | Unidentified species | 1 (2.0\%) |
| Apogon maculatus | 3 (6.1\%) | *Family Gammaridae* | 1 (2.0\%) |
| Phaeoptyx sp.* | 3 (6.1\%) | Unidentified species | 1 (2.0\%) |
| Family Blenniidae | 1 (2.0\%) | *Family Gonodactylidae* | 1 (2.0\%) |
| Parablennius marmoratus | 1 (2.0\%) | Neogonodactylus bredini* | 1 (2.0\%) |
| Family Callionymidae | 1 (2.0\%) | *Family Palaemonidae* | 1 (2.0\%) |
| Diplogrammus pauciradiatus* | 1 (2.0\%) | Unidentified species | 1 (2.0\%) |
| Family Carangidae | 3 (6.1\%) | *Family Penaeidae* | 3 (6.1\%) |
| Decapterus punctatus | 3 (6.1\%) | *Metapenaeopsis goodei* | 3 (6.1\%) |
| Family Chaenopodidae | 3 (6.1\%) | *Family Portunidae* | 1 (2.0\%) |
| Chaenopsis sp.* | 3 (6.1\%) | *Achelous ordwayi* | 1 (2.0\%) |
| Family Gobiidae | 11 (22.4\%) | Family Alpheidae | 1 (2.0\%) |
| Coryphopterus sp.* | 2 (4.1\%) | Unidentified species | 1 (2.0\%) |
| Microgobius sp.* | 10 (20.4\%) | *Family Gammaridae* | 1 (2.0\%) |
| Family Haemulidae | 5 (10.2\%) | *Family Palaemonidae* | 1 (2.0\%) |
| Haemulon aurolineatum | 5 (10.2\%) | | |
| Family Labridae | 2 (4.0\%) | Unidentified species | 1 (2.0\%) |
| Halichoeres sp.* | 1 (2.0\%) | | |
| Halichoeres bivittatus | 1 (2.0\%) | | |
| Family Microdesmidae | 12 (24.5\%) | *Family Alpheidae* | 1 (2.0\%) |
| Ptereleotris calliura | 12 (24.5\%) | Unidentified species | 1 (2.0\%) |
| Family Monacanthidae | 3 (6.1\%) | *Family Palaemonidae* | 1 (2.0\%) |
| Monacanthus ciliatus | 3 (6.1\%) | Unidentified species | 1 (2.0\%) |
| Family Pomacentridae | 5 (10.2\%) | *Achelous ordwayi* | 1 (2.0\%) |
| Chromis scotti | 3 (6.1\%) | | |
| Stegastes variabilis | 3 (6.1\%) | | |
| Family Scaridae | 1 (2.0\%) | | |
| Sparisoma atomarium* | 1 (2.0\%) | | |
| Family Serranidae | 8 (16.3\%) | *Family Alpheidae* | 1 (2.0\%) |
| Diplectrum formosum | 7 (14.3\%) | | |
| Hypoplectrus floridiae* | 1 (2.0\%) | | |
| Family Synodontidae | 8 (16.3\%) | *Family Palaemonidae* | 1 (2.0\%) |
| Synodus intermedius | 7 (14.3\%) | | |
| Synodus saurus* | 1 (2.0\%) | | |
DISCUSSION

Previous studies have shown lionfish in the Western Atlantic feed primarily on fish species and that a large degree of overlap in the taxonomic composition of prey exists across the region (Valdez-Moreno et al., 2012; Côté et al., 2013; Rocha et al., 2015; Harms-Tuohy, Schizas & Appeldoorn, 2016; Dahl et al., 2017, 2018; Sancho et al., 2018). By applying molecular approaches to identify prey fragments recovered from lionfish stomachs collected in the Gulf of Mexico off the South Florida coastline we report patterns broadly consistent with previous findings. Our results also demonstrate that adult lionfish in the study region preyed primarily on fish rather than invertebrates, as reflected in both the frequency in which each prey type was observed (i.e., 41 stomachs contained at least one fish prey item, while six stomachs had at least one invertebrate prey item) and the taxonomic richness of each prey type (e.g., 21 fish prey species compared to six crustacean species). Most prey species identified in this study were previously reported as lionfish prey in the Western Atlantic (Peake et al., 2018). The three higher taxa found in >20% of surveyed lionfish stomachs in this study were prevalent in lionfish stomachs in other areas of the Western Atlantic: with Apogonidae being prevalent in Puerto Rico (Harms-Tuohy, Schizas & Appeldoorn, 2016) and the northern Gulf of Mexico (Dahl et al., 2017), Gobiidae in Belize (Rocha et al., 2015), the Yucatan peninsula (Valdez-Moreno et al., 2012), the Bahamas (Côté et al., 2013), Biscayne National Park (Sancho et al., 2018), and the northern Gulf of Mexico (Dahl et al., 2017), and Microdesmidae in the Northern Gulf of Mexico (Dahl et al., 2017).

Such congruence in the taxonomic composition of lionfish prey amongst studies underscores that while generalists, these predators exhibit preferences for specific prey traits. Most fish prey species identified in this study share at least one of the following characteristics: they are reef-associated or demersal species with average sizes of <15 cm and are mostly active at night or during crepuscular times (Matheson et al., 2017). This is best exemplified by the most frequently observed prey taxa in this study, Ptereleotris calliura. This species was observed in ~25% of lionfish stomachs and is a small burrowing fish which does not exceed 12.5 cm in length (Robins & Ray, 1986). Similarly, Apogon species are small, nocturnal, reef associated fish (Hoese & Moore, 1998) and Microgobius are small reef-associated fish that may display burrowing behaviors (Birdsong, 1981). Diplectrum formosum while growing to relatively larger adult sizes than the above prey species, is a demersal species commonly associated with sandy and coarse gravel bottoms (Hoese & Moore, 1998). Lastly, Haemulon aurolineatum is a nocturnal reef-associated species that are <11 cm long in the juvenile stage (Collette & Talbot, 1972; Manooch & Barans, 1982). Such apparent preferences are in concordance with the findings of Green & Côté (2014), who using a trait-based model, identified nocturnal, small, shallow-bodied, solitary fishes found resting on or just above reefs as the most vulnerable to lionfish predation. Thus, our findings support the idea that lionfish, while generalists, exhibit a preference for small, nocturnal, reef associated fish prey in the Western Atlantic (see Peake et al. (2018) and references therein).
It is important to note that despite the overlap amongst results and past studies, we report patterns that suggest regional differences in lionfish diets may exist. We identified eleven putative novel lionfish prey taxa, nine species not identified in previous barcoding studies of lionfish stomach contents in the Western Atlantic [Diplogrammus pauciradiatus (Spotted dragonet), Halichoeres caudalis (Painted wrasse), Hypoplectrus floridae (Florida Hamlet), Phaeoptyx xenus (Sponge cardinalfish), Sparisoma atramentum (Greenblotch parrotfish), Synodus saurus (Atlantic lizardfish); Crustacean: Achelous ordwayi (Redhair swimming crab), Metapenaeopsis goodiei (Caribbean velvet shrimp), Neogonodactylus bredini (Caribbean rock mantis shrimp)] and two unidentified species from genera known to be lionfish prey (Chaenopsis sp. and Microgobius sp.). Based on observations reported in the Smithsonian Tropical Research Institute Online Information System (Robertson & Van Tassell, 2019) as well as the Global Biodiversity Information Facility databases (www.gbif.org), most of these novel prey taxa appear to be more prevalent in Florida waters than in those previously surveyed by barcoding studies. These findings indicate at least a partial effect of differences in regional assemblages in lionfish prey composition. Indeed, although comparisons across barcoding studies of lionfish in the Western Atlantic are hampered by the application of different methodologies and the reporting of different frequency statistics, both prey composition and frequencies observed in our study appear to be the most similar to those reported by Dahl et al. (2017) from nearby Northern Gulf of Mexico waters (see Tables S1 and S2). Lastly, most prey species reported in this study are species frequently to commonly found in reef-assemblages in Florida waters (Smith et al., 1975; Darcy & Guthertz, 1984; Matheson et al., 2017). These patterns thus give credence to the Harms-Tuohy, Schizas & Appeldoorn (2016) suggestion that lionfish diets are a result of regional ichthyofauna and their preferences for specific traits. As such, continued research is important to elucidate lionfish diets in unstudied regions and habitats to understand local impacts on species and communities.

Similarly, future studies should investigate whether lionfish of various size classes prey upon different taxa. Although ontogeny appears unlikely to have affected results reported herein given the size distribution of lionfish sampled, ontogenetic shifts in diet are known to occur in lionfish with studies indicating a shift from an invertebrate-to a fish dominated diet with increase in size (Morris & Akins, 2009; Muñoz, Currin & Whitfield, 2011; Dahl & Patterson, 2014; O’Farrell et al., 2014; Peake et al., 2018; Sancho et al., 2018; Malpica-Cruz, Green & Côté, 2019). For instance, Peake et al. (2018) reported significant negative and positive relationships between lionfish size (as measured by standard length) and the contributions of shrimp and fish prey respectively to lionfish diets in the Arrecifes de Cozumel National Park. Similarly, Sancho et al. (2018) reported that in Biscayne Bay National Park, lionfish sized <180 mm TL exhibited diets dominated by shrimp prey as measured by Index of Relative Importance and those larger than 180 mm TL exhibited diets dominated by fish. Sancho et al. (2018) also reported some differences in prey composition across size classes (e.g., pleocyemate shrimp only preyed by smaller lionfish, dendrobranchiate shrimp by larger classes); however, the extent to which ontogeny affects the composition of lionfish prey remains unclear. Future barcoding studies may aid in elucidating this point as well as the drivers of any such changes.
CONCLUSION

In this study, we provide the first report of the lionfish prey-items from the Florida coastlines of the Gulf of Mexico. Our results are largely congruent with previous findings that suggest lionfish are generalists whose diet reflects regional species compositions and prey preferences for specific behavioral and morphological traits. Despite a limited number of samples and collection period, the results of our study were quite similar to the taxonomic composition of prey items reported by Dahl et al. (2017) from the Northern Gulf of Mexico, and suggest the former to be the most likely explanation. We report eleven putative novel prey species of lionfish, possibly reflecting the local prey availability in the habitat where lionfish were collected. We suggest molecular based efforts to understand the dietary habits of lionfish throughout their non-native region continue, with lionfish derbies serving as a cost-effective way to leverage public events into producing scientific information.

ACKNOWLEDGEMENTS

We thank the organizers and participants of the 2016 Sarasota Lionfish Derby for collecting lionfish used in this study.

ADDITIONAL INFORMATION AND DECLARATIONS

Funding

Funding for this work was provided by the University of South Florida Sarasota-Manatee and Mote Marine Laboratory. The funders had no role in study design, data collection and analysis, decision to publish, or preparation of the manuscript.

Grant Disclosures

The following grant information was disclosed by the authors:
University of South Florida Sarasota-Manatee.
Mote Marine Laboratory.

Competing Interests

The authors declare that they have no competing interests.

Author Contributions

• Carlos A. Santamaria conceived and designed the experiments, performed the experiments, analyzed the data, prepared figures and/or tables, authored or reviewed drafts of the paper, and approved the final draft.
• James Locascio conceived and designed the experiments, performed the experiments, authored or reviewed drafts of the paper, and approved the final draft.
• Taylor M. Greenan conceived and designed the experiments, performed the experiments, analyzed the data, authored or reviewed drafts of the paper, and approved the final draft.
Animal Ethics
The following information was supplied relating to ethical approvals (i.e., approving body and any reference numbers):

Capture and euthanasia of lionfish was not performed by authors; however, all participants of the Sarasota Lionfish Derby in which lionfish were collected followed all local, state, and federal guidelines and stipulations. These include those specified by the Florida Fish and Wildlife Conservation Commission (see https://myfwc.com/fishing/saltwater/recreational/lionfish/).

Data Availability
The following information was supplied regarding data availability:

Raw data is available as a Supplemental File.

Supplemental Information
Supplemental information for this article can be found online at http://dx.doi.org/10.7717/peerj.9922#supplemental-information.

REFERENCES

Aguilar-Perera A, Tuz-Sulub A. 2010. Non-native, invasive red lionfish (Pterois volitans [Linnaeus, 1758]: Scorpaenidae), is first recorded in the southern Gulf of Mexico, off the northern Yucatan Peninsula. Mexico Aquatic Invasions 5(Suppl. 1):S9–S12 DOI 10.3391/ai.2010.5.S1.003.

Albins MA. 2013. Effects of invasive Pacific red lionfish Pterois volitans versus a native predator on Bahamian coral-reef fish communities. Biological Invasions 15(1):29–43 DOI 10.1007/s10530-012-0266-1.

Albins MA. 2015. Invasive Pacific lionfish Pterois volitans reduce abundance and species richness of native Bahamian coral-reef fishes. Marine Ecology Progress Series 522:231–243 DOI 10.3354/meps11159.

Albins MA, Hixon MA. 2008. Invasive Indo-Pacific lionfish Pterois volitans reduce recruitment of Atlantic coral-reef fishes. Marine Ecology Progress Series 367:233–238 DOI 10.3354/meps07620.

Birdsong RS. 1981. A review of the gobiid fish genus Microgobius Poey. Bulletin of Marine Science 31:267–306.

Bogdanoff AK, Mostowy J, Peake J, Layman CA, Bermudez AB, Baca CG, Palacios NH, Gonzalez DTM, Xicoténcatl MDRB, Morris JA Jr. 2018. A brief description of invasive lionfish (Pterois sp.) diet composition in the Arrecifes de Cozumel National Park. Food Webs 17:e00104 DOI 10.1016/j.foodweb.2018.e00104.

Chang C-H, Shao K-T, Lin H-Y, Chiu Y-C, Lee M-Y, Liu S-H, Lin P-L. 2017. DNA barcodes of the native ray-finned fishes in Taiwan. Molecular Ecology Resources 17(4):796–805 DOI 10.1111/1755-0998.12601.

Collette BB, Talbot FH. 1972. Activity patterns of coral reef fishes with emphasis on nocturnal-diurnal changeover. In: Collette BB, Earle SA, eds. Results of the Tektite Program: Ecology of Coral Reef Fishes. Los Angeles County: National Museum of Natural History, 98–124.

Côté IM, Green SJ, Morris JA Jr, Akins JL, Steinke D. 2013. Diet richness of invasive Indo-Pacific lionfish revealed by DNA barcoding. Marine Ecology Progress Series 472:249–256 DOI 10.3354/meps09992.
Dahl KA, Patterson WF III. 2014. Habitat-specific density and diet of rapidly expanding invasive red lionfish, *Pterois volitans*, populations in the northern Gulf of Mexico. *PLOS ONE* 9(8):e105852 DOI 10.1371/journal.pone.0105852.

Dahl KA, Patterson WF, Robertson A, Ortmann AC. 2017. DNA barcoding significantly improves resolution of invasive lionfish diet in the Northern Gulf of Mexico. *Biological Invasions* 19(6):1917–1933 DOI 10.1007/s10530-017-1407-3.

Dahl KA, Portnoy DS, Hogan JD, Johnson JE, Gold JR, Patterson WF. 2018. Genotyping confirms significant cannibalism in northern Gulf of Mexico invasive red lionfish, *Pterois volitans*. *Biological Invasions* 20(12):3513–3526 DOI 10.1007/s10530-018-1791-3.

Darcy GH, Gutherz EJ. 1984. Abundance and density of demersal fishes on the West Florida Shelf. *Bulletin of Marine Science* 34:81–105.

Folmer O, Black M, Hoeh W, Lutz R, Vrijenhoek R. 1994. DNA primers for amplification of mitochondrial cytochrome c oxidase subunit I from diverse metazoan invertebrates. *Molecular Marine Biology and Biotechnology* 3:294–299.

Green SJ, Akins JL, Côté IM. 2011. Foraging behaviour and prey consumption in the Indo-Pacific lionfish on Bahamian coral reefs. *Marine Ecology Progress Series* 433:159–167 DOI 10.3354/meps09208.

Green SJ, Akins JL, Maljković A, Côté IM. 2012. Invasive Lionfish drive Atlantic coral reef fish declines. *PLOS ONE* 7(3):e32596 DOI 10.1371/journal.pone.0032596.

Green SJ, Akins JL, Morris JA Jr. 2012. Lionfish dissection: techniques and applications. NOAA Technical Memorandum NOS NCCOS 139. Available at https://www.reef.org/sites/default/files/_LF_dissection_final.pdf.

Green SJ, Côté IM. 2014. Trait-based diet selection: prey behaviour and morphology predict vulnerability to predation in reef fish communities. *Journal of Animal Ecology* 83(6):1451–1460 DOI 10.1111/1365-2656.12250.

Harms-Tuohy CA, Schizas NV, Appeldoorn RS. 2016. Use of DNA metabarcoding for stomach content analysis in the invasive lionfish *Pterois volitans* in Puerto Rico. *Marine Ecology Progress Series* 558:181–191 DOI 10.3354/meps11738.

Hoese HD, Moore RH. 1998. *Fishes of the Gulf of Mexico, Texas, Louisiana, and adjacent waters*. Second Edition. College Station: Texas A&M University Press.

Ingeman KE. 2016. Lionfish cause increased mortality rates and drive local extirpation of native prey. *Marine Ecology Progress Series* 558:235–245 DOI 10.3354/meps11821.

Knebelsberger T, Landi M, Neumann H, Kloppmann M, Sell AF, Campbell PD, Laakmann S, Raupach MJ, Carvalho GR, Costa FO. 2014. A reliable DNA barcode reference library for the identification of the North European shelf fish fauna. *Molecular Ecology Resources* 14:1060–1071.

Leray M, Yang JY, Meyer CP, Mills SC, Agudelo N, Ranwez V, Boehm JT, Machida RJ. 2013. A new versatile primer set targeting a short fragment of the mitochondrial COI region for metabarcoding metazoan diversity: application for characterizing coral reef fish gut contents. *Frontiers in Zoology* 10(1):34 DOI 10.1186/1742-9994-10-34.

Malpica-Cruz L, Green SJ, Côté IM. 2019. Temporal and ontogenetic changes in the trophic signature of an invasive marine predator. *Hydrobiologia* 839(1):71–86 DOI 10.1007/s10750-019-03996-2.

Manooch CS, Barans CA. 1982. Distribution, abundance, and age and growth of the Tomtate, *Haemulon aurolineatum*, along the southeastern United States coast. *Fishery Bulletin* 80:1–19.
Matheson RE Jr, Flaherty-Walia KE, Switzer TS, McMichael RH Jr. 2017. The importance of time of day in structuring demersal ichthyofaunal assemblages on the West Florida Shelf. *Bulletin of Marine Science* 93(2):407–437 DOI 10.5343/bms.2016.1047.

Mizrahi M, Chapman JK, Gough CLA, Humber F, Anderson LG. 2017. Management implications of the influence of biological variability of invasive lionfish diet in Belize. *Management of Biological Invasions* 8(1):61–70 DOI 10.3391/mbi.2017.8.1.06.

Morris JA, Akins JL. 2009. Feeding ecology of invasive lionfish (*Pterois volitans*) in the Bahamian archipelago. *Environmental Biology of Fishes* 86(3):389–398 DOI 10.1007/s10641-009-9538-8.

Muñoz RC, Currin CA, Whitfield PE. 2011. Diet of invasive lionfish on hard bottom reefs of the Southeast USA: insights from stomach contents and stable isotopes. *Marine Ecology Progress Series* 432:181–193 DOI 10.3354/meps09154.

O’Farrell S, Bearhop S, McGill RAR, Dalgrogen CP, Brumbaugh DR, Mumby PJ. 2014. Habitat and body size effects on the isotopic niche space of invasive lionfish and endangered Nassau grouper. *Ecosphere* 5(10):1–11 DOI 10.1890/ES14-00126.1.

Peake J, Bogdanoff AK, Layman CA, Castillo B, Reale-Munroe K, Chapman J, Dahl K, Patterson WF, Eddy C, Ellis RD, Faletti M, Higgs N, Johnston MA, Muñoz RC, Sandel V, Villasenor-Derbez JC, Morris JA Jr. 2018. Feeding ecology of invasive lionfish (*Pterois volitans* and *Pterois miles*) in the temperate and tropical western Atlantic. *Biological Invasions* 20(9):2567–2597 DOI 10.1007/s10530-018-1720-5.

Ratnasingham S, Hebert PDN. 2007. Bold: the barcode of life data system (http://www.barcodinglife.org). *Molecular Ecology Notes* 7(3):355–364 DOI 10.1111/j.1471-8286.2007.01678.x.

Robertson DR, Van Tassell J. 2019. Shorefishes of the Greater Caribbean: online information system. Version 2.0. Balboa: Smithsonian Tropical Research Institute.

Robins CR, Ray GC. 1986. *A field guide to Atlantic coast fishes of North America*. Boston: Houghton Mifflin Company.

Rocha LA, Rocha CR, Baldwin CC, Weigt LA, McField M. 2015. Invasive lionfish preying on critically endangered reef fish. *Coral Reefs* 34(3):803–806 DOI 10.1007/s00338-015-1293-z.

Sancho G, Kingsley-Smith PR, Morris JA, Toline CA, McDonough Y, Doty SM. 2018. Invasive lionfish (*Pterois volitans*/*miles*) feeding ecology in Biscayne National Park. *Biological Invasions* 20(9):2343–2361 DOI 10.1007/s10530-018-1705-4.

Schofield PJ. 2010. Update on geographic spread of invasive lionfishes (*Pterois volitans* [Linnaeus, 1758] and *P. miles* [Bennett, 1828]) in the Western North Atlantic Ocean, Caribbean Sea and Gulf of Mexico. *Aquatic Invasions* 5(Suppl. 1):S117–S122 DOI 10.3391/ai.2010.5.S1.024.

Smith GB, Austin HM, Bortone SA, Hastings RW, Ogren LH. 1975. *Fishes of the Florida Middle Ground with comments on ecology and zoogeography*. Page 14 *Florida Marine Research Publications*. St. Petersburg: Florida Department of Natural Resources, Marine Research Laboratory.

Valdez-Moreno M, Quintal-Lizama C, Gómez-Lozano R, García-Rivas MDC. 2012. Monitoring an alien invasion: DNA barcoding and the identification of lionfish and their prey on coral reefs of the Mexican Caribbean. *PLOS ONE* 7(6):e56636 DOI 10.1371/journal.pone.0056636.

Zhang J, Hanner R. 2012. Molecular approach to the identification of fish in the South China Sea. *PLOS ONE* 7(2):e30621 DOI 10.1371/journal.pone.0030621.