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Examining the status of improved air quality in world cities due to COVID-19 led temporary reduction in anthropogenic emissions

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1. Introduction

As per the Ecosystem Services (ESs) definition of Millennium Ecosystem Assessment (MA, 2005), clean air is one of the fundamental needs of human lives (Ash et al., 2010; Baró et al., 2014; Schirpke et al., 2014; Charles et al., 2020). Air pollution has been reduced substantially during the COVID-19 lockdown period. Venter et al. (2020) had examined both tropospheric and ground air pollution levels using satellite data and a network of >10,000 air quality stations across the world and found 29% reduction in NO2 (with 95% confidence interval – 44% to –13%), 11% reduction in Ozone (O3), and 9% reduction in PM2.5 during the first two weeks of lockdown (Venter et al., 2020). Kerimray et al.
which was ~53% at urban stations. During the same period, the mean concentration of five key air pollutants, i.e., SO$_2$, PM$_{2.5}$, PM$_{10}$, NO$_2$, and CO have decreased by ~7%, ~6%, ~14%, ~25%, and ~5%, respectively (Kerimray et al., 2020). Mahato et al. (2020) had reported a sharp reduction in air pollution in Delhi, one of the most polluted cities in the world. The author found that the concentration of PM$_{10}$ and PM$_{2.5}$ in Delhi was reduced to 60% and 39%, compared to the air pollution levels in 2019. The concentration of other pollutants, such as NO$_2$ (~53%) and CO (~30%), have also been reduced substantially during the lockdown period. In addition to this, Mahato et al. study had observed a 40%–50% improvement in air quality in Delhi within the first week of lockdown. Bao and Zhang (2020) study combined air pollution and Intracity Migration Index (IMI) data for 44 cities in northern China and found that restriction on human mobility is strongly associated with the reduction of air pollution in these cities. The author found that the air quality index (AQI) in these cities is decreased by ~8%, as the concentration of five key air pollutants, i.e., SO$_2$, PM$_{2.5}$, PM$_{10}$, NO$_2$, and CO have decreased by ~7%, ~6%, ~14%, ~25%, and ~5%, respectively (Sicard et al., 2020) had observed that due to COVID-19 lockdown, NO$_2$ mean concentrations were reduced substantially in all European cities, which was ~53% at urban stations. During the same period, the mean concentrations of O$_3$ was reported to be increased at the urban stations in Europe, i.e., 24% increases in Nice, 14% increases in Rome, 27% increases in Turin, 2.4% increases in Valencia and 36% in increases in Wuhan (China). Otmani et al. (2020) study at Morocco using three-dimensional air mass backward trajectories and the HYSSPLIT model found that PM$_{10}$, SO$_2$, and NO$_2$ have been reduced up to 75%, 49%, and 96% during the lockdown period. In the southeast Asian (SEA) countries, Kanniah et al. (2020) study found that in Malaysia, PM$_{10}$, PM$_{2.5}$, NO$_2$, SO$_2$, and CO concentrations have been decreased by 26–31%, 23–32%, 63–64%, 9–20%, and 25–31% during the lockdown period. Kumar et al. (2020a) examined the impacts of COVID-19 mitigation measures on the reduction of PM$_{2.5}$ in five Indian cities (Chennai, Delhi, Hyderabad, Kolkata, and Mumbai), using in-situ measurements from 2015 to 2020, and termed it as an ‘anthropogenic emission switch-off’ experiment, allowing to understand the baseline concentrations across various cities. Kumar et al. study found that during the lockdown period (25 March to 11 May), the PM$_{2.5}$ concentration in the selected cities has been reduced by 19–43% (Chennai), 41–53% (Delhi), 26–54% (Hyderabad), 24–36% (Kolkata), and 10–39% (Mumbai), respectively. This study also found that cities with higher traffic volume exhibited a more significant reduction of PM$_{2.5}$. The level of air pollution has a severe impact on human health and overall well-being. Air pollution is responsible for nearly 5 million deaths each year globally (IHME 2018). In 2017, air pollution had contributed to 9% of total deaths, ranges from 2% in the highly developed country to a maximum of 15% in low-developed countries, especially in South and East Asia (IHME 2018). Based on Disability-Adjusted Life Years (DALYs) statistics, which demonstrate of losing one year of good health due to either premature mortality or disability caused by any factors, it has been estimated that air pollution is the 5th largest contributor to overall disease burden, only after high blood pressure, smoking, high blood sugar, and obesity, respectively (IHME 2018). The adverse impact of air pollution on human health is not only limited to low-developed countries. In the European regions, nearly 193,000 deaths in 2012 were attributed to airborne particulate matter (Critz et al., 2017). In addition, it has been found that air pollution in China is accountable for 4000 deaths each day, i.e., 1.6 million casualties in 2016 (Wang and Hao, 2012; Rohde and Muller, 2015). Chen et al. (2020) found that reduction in PM$_{2.5}$ during the lockdown period helped to avoid a total of 3214 PM$_{2.5}$ related deaths (95% CI 2340–4087). Chen et al. (2020) also estimated that COVID-19 lockdown and resulted cut down of air pollution brought multi-faceted health benefits to non-COVID mortalities. Several research studies (Crouse et al., 2015; Dutheil et al., 2020a; He et al., 2020) have echoed the surmountable effects of air pollutants on human lives and found that an increase in 10 µg/m$^3$ of NO$_2$ per day will be responsible for a 0.13% increases of all-cause mortality (He et al., 2020). The mortality rate would be around 2% when the 5-day average NO$_2$ level would reach 10 µg/m$^3$ (Monica et al., 2011).

It is now well-established by many data-driven experiments that the accelerated rate of air pollution can have a substantial impact on overall human well-being. Many previous studies have examined the synergistic association between limited human activity and improved air quality across the scale (Chen et al., 2020; Kumar et al., 2020; Mahato et al., 2020; Ogen, 2020). These studies collectively suggested that temporary or periodic cessation of human activity could be a temporary solution for battling air pollution. However, the substantial reduction in air pollution during the lockdown is obvious and does not convey any revelation. Therefore, the co-benefits of this reduced anthropogenic emission need to be evaluated with an evidence-based approach to allow for the results to be used as a reference for future decision making and policy development. In addition to this, most of the previous studies have relied on ground-based measurements, and hence, strongly depend on the availability of publicly available data, which often creates obstacles while upscaling the approaches for larger scales. Therefore, this research work has made an effort to assess the air pollution levels of many key air pollutants after combining both satellite and ground measurements. This work aims to estimate the spatiotemporal variations in air pollution levels during the lockdown period from 1 February to 11 May in 2020 using a reference of the same period in 2019. The avoided premature mortality due to the reduction of air pollution levels and the corresponding economic benefits were also assessed using multiple economic valuation approaches. Finally, each city’s population-weighted average air pollution concentration was estimated for the considered study period.

2. Materials and methods

2.1. Data source and data preparation

A total of 20 cities have been selected for evaluating the effect of lockdown on air quality. These cities are Antwerp, Barcelona, Brussels, Chicago, Cologne, Denver, Frankfurt, London, Los Angeles, Madrid, Milan, New York, Paris, Philadelphia, Rotterdam, Sao Paulo, Tehran, Turin, and Utrecht. These cities have been considered based on two criteria: high air pollution and high COVID-19 casualties. Most of the cities listed here are from European and American countries. These countries reported more COVID-19 casualties compared with the Asian and Latin American countries (as of May 11, 2020) (Sannigrahi et al., 2020; WHO, 2020). Both satellite remote sensing and ground air pollution data were utilized for evaluating the positive effects of lockdown on the air quality levels of these cities. For comparison purposes, the satellite-based air pollution was measured from 1 February to 11 May for both 2019 (lockdown equivalent period) and 2020 (lockdown period). The concentration of two key air pollutants, nitrogen dioxide (NO$_2$) and carbon monoxide (CO) was computed for both 2019 and 2020 using Sentinel 5 P data. Human mobility data Our World in Data including driving and transit for the selected cities, were collected from Apple (https://covid19.apple.com/mobility/) (city-scale) and Google (https://www.google.com/covid19/mobility/) (country-scale) mobility reports. In addition, the gridded human settlement data and population density data (pixel format) were collected from the Socio-Economic Data Application Center, National Aeronautics and Space Application data center (SEDAC, NASA). For evaluating the total air pollution reduction of these 20 cities in a more accurate way, the Geographical Information System (GIS) enabled city boundary (shapefile format) was extracted from the OpenStreetMap (OSM) application. Two consecutive steps were followed to get the boundary of these cities. First, the OSM relation identifier number (OSM id) was generated for all the 20 cities using
2.2. Estimation of air pollution

2.2.1. Sentinel 5 P TROPOMI data and TROPOMI Explorer Application

The ESA (European Space Agency) Sentinel-5 Precursor (S 5 P) is an example of low earth Sun-synchronous Orbit (SSO) polar satellite that provides information of tropospheric air quality, climate dynamics, and ozone layer concentration for the time period 2015–2022 (Veehkind et al., 2012). The ESA S 5 P mission is one of the few missions that is intended to measure air and climatic variability from the space-borne application. The S 5 P mission is associated with the Global Monitoring of the Environment and Security (GMES) space programme. The TROPOspheric Monitoring Instrument (TROPOMI) payload of S 5 P mission was designed to measure the tropospheric concentration of few key air pollutants, i.e., Ozone (O$_3$), NO$_2$, SO$_2$, CO, CH$_4$, CH$_2$O, and aerosol properties in line with Ozone Monitoring Instrument (OMI) and SCanning Imaging Absorption spectroMeter for Atmospheric CartograpHY (SCIAMACHY) programme (Veehkind et al., 2012). TROPOMI measures the concentration of key tropospheric constituents at a 7 × 3.5 km$^2$ spatial unit. This default spatial scale was downscaled into 1 km × 1 km scale for city-scale analysis and subsequent interpretation. In this study, the spatial and temporal variability of two key air pollutants was extracted and mapped from the TROPOMI measurements using the Google Earth Engine cloud platform. For this purpose, an interactive application called TROPOMI Explorer Application (https://showcase.earthengine.app/view/tropomi-explorer), was utilized to facilitate quick and easy SSP data exploration and to examine the changes in air pollution in both cross-sectional and longitudinal scale. Spatial visualization and time series charts for the selected air pollutants were prepared with the help of the TROPOMI Explorer application. The other accessories of this application, such as NO$_2$ time series inspector, NO$_2$ temporal comparison, NO$_2$ time-series animation, were also utilized for different computational purposes.

An extensive body of research has examined the applicability of satellite remote sensing in air pollution assessment from regional to global scale (Meng et al., 2016; Fernández-Pacheco et al., 2018; Alvar- ez-Mendoza et al., 2019; Basu et al., 2019; Zhang et al., 2019). Many studies have focused on monitoring Aerosol Optical Depth (AOD) using Moderate Resolution Imaging Spectroradiometer (MODIS) data to predict ambient fine particulate matter concentration (Zhang et al., 2010; Mehta et al., 2016). Since MODIS data has a low spatial resolution that often limits its application at fine-scale air pollution assessment, several moderate to high-resolution satellite data products, such as Landsat and Sentinel emerged to be an efficient alternative to measure pollution levels at city scale (Meng et al., 2016; Basu et al., 2019). Several approaches, including land use regression (Kloog et al., 2012; Basu et al., 2019), machine learning approaches – random forest (Fernández-Pacheco et al., 2018), stepwise regression, partial least squares regression (Alvarez-Mendoza et al., 2019), artificial neuronal network (Zhang et al., 2019), have been established to retrieve AOD and particulate matter concentration from Landsat and Sentinel data.

2.2.2. In-situ air pollution data

Ground monitored air quality data was collected from different governmental sources and open data repositories, including U.S. Environmental Protection Agency (https://www.epa.gov) (for Chicago, Denver, Detroit, Los Angeles, New York, and Philadelphia), European Environmental Agency (https://www.eea.europa.eu/) (for Antwerp, Barcelona, Brussels, Frankfurt, London, Madrid, Milan, Paris, Rotterdam, Utrecht), and OpenAQ (https://openaq.org/) (for Detroit and Los Angeles). The in-situ data was collected for three key air pollutants, i.e., NO$_2$, PM$_{2.5}$, and PM$_{10}$, for a fixed time period (1 February to 11 May) of both 2019 and 2020. The average concentration of different air pollutants was calculated to perform comparative assessment and subsequent interpretation. Since the in-situ air pollution data was not adequate for thorough spatial assessment, the same had not been used for validating satellite pollution estimates. The time series (2000–2020) air quality index (AQI) of the US cities (time-series historical data is not available for other cities) was also generated using the multilayer time plot function of EPA. The overall AQI values were sub-divided into six groups, i.e., good, moderate, unhealthy for sensitive population groups, unhealthy, very unhealthy, and hazardous, respectively. In addition, the single year AQI data was also extracted for the selected cities from the EPA. The number of unhealthy days for each pollutant was measured using the EPA AQI plot function. The combination of two different pollutants, such as CO and NO$_2$, PM$_{10}$, and PM$_{2.5}$, was used to assess the yearly AQI status of the cities. As several studies reported the increment of O$_3$ due to the reduction of GHG emissions, this study also evaluated the O$_3$ exceedances for the current year compared to the average O$_3$ concentration of the last 5 and 20 years. This particular task was implemented using the EPA Ozone exceedances plot function (EPA, 2020). Table S1 provides the criteria of categorization for each index.

2.3. Environmental significance of improving air quality status

The accelerating increases of air pollution in cities is a major concern across the world (Mayer, 1999; Kim Oanh et al., 2006; Chan and Yao, 2008; Guttikunda et al., 2014; Pilla and Broderick, 2015; Abhiijith et al., 2017; Rai et al., 2017; Zhu et al., 2020; Kumar et al., 2021; Rodriguez-Urrego and Rodriguez-Urrego, 2020). Various policies have been implemented for managing the city-based air pollution that mainly originated from anthropogenic activities from specific sources and sectors (Baró et al., 2014; Kumar et al., 2015, 2016, 2019a; Feng and Liao, 2016; Zhang et al., 2016). These include the Directive 2010/75/EU on industrial emissions, initiated by European Commission to define “Euro standards” for measuring the road vehicle emissions and the Directive 94/63/EC for calculating volatile organic compounds emissions from petrol storage (Baró et al., 2014). The reduction of these gaseous pollutants by green canopies has significant economic importance (Kumar et al., 2019). Two main ecosystem services, such as air quality regulation and climate/gas regulation, are mainly associated with air quality ecosystem services (Zhang et al., 2018, 2020). Several studies have calculated the economic values of NO$_2$, SO$_2$, CO reductions using various valuation approaches such as carbon tax, the social cost of carbon, shadow price method, marginal cost method. (Guerriero et al., 2016; Castro et al., 2017; Jeanjean et al., 2017; Bhervani et al., 2020). Since this study has considered the air pollution reduction at the city scale, the public health burden (utilized for in-situ data-based economic valuation) and mean externality valuation (utilized for satellite data-based economic valuation) approaches were utilized for estimating economic damage due to air pollution and for calculating the economic values of improved air quality (Matthews and Lave, 2000; Baro et al., 2014). Unit social damage price due to air pollution was estimated for 2020 using the US consumer price index (CPI) inflation calculator (U.S. Bureau of Labor Statistics, 2020). Additionally, using the most updated price conversion factors, the mean externality values for the key pollutants were estimated as 5149 and 956 US$ ton$^{-1}$ for NO$_2$ and CO, respectively.

The public health burden valuation approach has been utilized in many studies for health impact assessment (COMEAP, 2010; Hu et al., 2015; Sahi and Kota, 2017; Etchie et al., 2018; Kumar et al., 2020b; Sharma et al., 2020). The calculation of public health burden and the associated economic burden was conducted by following three subsequent steps: first, estimation of population-weighted average concentration; second, estimation of health burden or a number of premature
mortality attributable to air pollution; and third, the economic burden due to excess air pollution and economic benefits subject to the reduction of air pollution levels during the lockdown period. The population-weighted average concentration (PWAC) (Ivy et al., 2008; Etchie et al., 2018; Park et al., 2020) was measured as follows:

$$PWAC = \frac{\sum x (Pop_x \times C_x)}{\sum Pop_x}$$  \hspace{1cm} (1)

where \(Pop_x\) is the population count of a pixel, \(C_x\) is the average concentration of NO\(_2\), PM\(_{2.5}\), PM\(_{10}\) (101 days, 1 February to 11 May in 2019 and 2020). \(\sum Pop_x\) is the total population count of the city, PWAC(\(\mu g/m^3\)) is the population-weighted average concentration. The PWAC was estimated using the ArcPy Python module. Gridded population data from SEDAC, NASA was utilized for this task. Pollution and gridded population data for the same time period were used for estimations of PWAC.

Following, the health burden (HB), which refers to premature deaths attributable to short-term exposure to air pollutants, was estimated for the study period. The reduction in HB (\(\Delta HB\)) was also measured by calculating the difference between the previous and later HB estimates.

$$HB_x = AFI_x \times B_x \times \sum Pop_x$$  \hspace{1cm} (2)

$$AF = \left( \frac{R_{i,2020} - 1}{R_{i,2019}} \right)$$  \hspace{1cm} (3)

$$\Delta HB = HB_{2019} - HB_{2020}$$  \hspace{1cm} (4)

$$RR_i = e^{[\beta_i (C_i - C_0)]}$$. \(C > 0\)  \hspace{1cm} (5)

$$ER = RR - 1$$  \hspace{1cm} (6)

where \(HB_x\) is the health burden of city \(x\), \(AF\) is the attributable fraction associated with the relative risk of each pollutant, \(RR_i\) is the relative risk of pollutant \(i\), \(B_x\) is the baseline-covariate mortality rate per 100,000 population. For calculating \(B_x\), the country-wise cardiovascular and chronic respiratory baseline mortality rate was collected from the Global Burden of Disease study of 2017 (IHME, 2018). \(Pop_x\) is the population of city \(x\) derived from the SEDAC, NASA gridded population count data. \(\Delta HB\) is the difference in health burden (or avoidance of premature death due to the reduction in air pollution) from 1 February to May 11, 2020 compared to the same period in 2019. \(HB_{2019}\) and \(HB_{2020}\) is the health burden estimates in 2019 and 2020 (estimated for 1 February to 11 May time period). \(\beta\) is the exposure-response relationship coefficient, indicates the excess risk of health burden (such as mortality) per unit increase of pollutants. \(\beta\) is calculated 0.038%, 0.032%, 0.081%, 0.13%, and 0.048% per 1 \(\mu g/m^3\) increases of PM\(_{2.5}\), PM\(_{10}\), SO\(_2\), NO\(_2\), and O\(_3\), respectively (Hu et al., 2015; Chen et al., 2020; Kumar et al., 2020a; Sharma et al., 2020). \(\beta\) is calculated 3.7% per 1 mg/m\(^3\) increase of CO. \(C_0\) is the concentration of pollutant \(i\), \(C_i\) is the threshold concentration, below which the pollutant exhibits no obvious adverse health effects (i.e., RR = 1).

The economic burden (EB) and economic benefits of the reduced air pollution concentration were estimated using the value of statistical life (VSL) approach (Hu et al., 2015; Etchie et al., 2018). The VSL represents an individual’s willingness to pay for a marginal reduction in the risk of dying. The economic benefits due to avoided premature mortality were estimated as follows:

$$EB = HB \times VSL$$  \hspace{1cm} (7)

where \(EB\) is the economic benefit attributed to the reduction of air pollution and resulted in estimates of avoidable mortality, \(HB\) is the health burden estimates of city \(x\), VSL is the value of statistical life of the country \(x\) that corresponds to the city. Since this study considers cities that cover many diversified economic setup and development background, a uniform income elastic global VSL estimates measured by Viscusi et al. (2017) was considered for the economic valuation and subsequent analysis. As city-specific VSL data is not available for many cities, the VSL estimates for the corresponding countries were taken for the analysis. The 2017 VSL values were converted to 2020 unit price for adjusting price fluctuation. The income adjusted VSL (Million US$) was estimated as Belgium (8, was used for Antwerp and Brussels), Spain (5, was used for Barcelona, Madrid), the USA (10, was used for Chicago, Denver, Detroit, Los Angeles, New York, and Philadelphia), Germany (8, was used for Cologne, Frankfurt), the UK (8, was used for London), Italy (6, was used for Milan and Turin), France (7, was used for Paris), the Netherlands (9, was used for Rotterdam and Utrecht), Brazil (2, was used for Sao Paulo), and Iran (1, was used for Tehran), respectively (Viscusi et al., 2017).

### 2.4. Examining human mobility and its connections with air pollution status

Due to the emergence of the COVID-19 pandemic, countries across the world imposed mandatory lockdowns to restrict human-mobility. This triggers the reduction of motorized traffic, which is one of the key sources of urban air pollution (Chinazzi et al., 2020; De Brouwer et al., 2020). Human mobility could accelerate the transmission of contagious diseases, especially when a larger fraction of the population daily commutes used public transport to sustain their essential daily journey (Sasidharan et al., 2020). Troko et al. (2011) study noted a statistically significant association between human mobility that is mainly attributed to public transport and transmissions of acute respiratory infections (ARI). Troko et al. (2011) also found that the use of public transport within the five days of symptom onset (Influenza) in the UK has increased the risk of ARI infection by six-times. To evaluate the effects of reduced human mobility on air pollution, the study presented in this paper utilized the human mobility data provided by Apple and Google (Drake et al., 2020; Wang and Yamamoto, 2020; Wellenius et al., 2020; Yilmazkuday, 2020). Apple mobility data includes three mobility components, i.e., driving, walking, and transit (public transport), respectively. The reduction of human mobility during the lockdown period was calculated from the baseline (13 January). Both positive and negative changes in human mobility were recorded in percentage form to eliminate calculation bias. Among the three mobility components, driving and transit was considered for the evaluation. Google mobility data was also used in this study which has six components (retail and recreation, grocery and pharmacy, parks, transits, workplace, and residential). This data is available from February 15, 2020 to recent date. Since Google mobility data is not available for city scale, the smallest scale (county/state) was taken for the analysis for which the mobility counts are available. This data is also prepared in percentage format to handle the calculation bias.

### 3. Results

#### 3.1. Spatiotemporal changes in air pollution in different cities

Spatial distribution of tropospheric NO\(_2\) and CO column (derived from Sentinel TROPOMI data) is analyzed and presented in Fig. 1 and Fig. S1. During the observation period, a sharp reduction in NO\(_2\) and CO (\(\mu g/m^3\)) is observed in all 20 cities. This could be due to the lockdown and resultant reduction of transportation and industrial emission. The maximum reduction in NO\(_2\) is found for the European cities, such as Paris, Milan, Madrid, Turin, London, Frankfurt, Cologne, and American cities, such as New York, Philadelphia, etc. (Fig. 1). Moreover, the highest and lowest NO\(_2\) reduction is found in Tehran and Sao Paulo. The CO concentration has also been reduced significantly during the study period. The highest reduction is recorded in Detroit, followed by Barcelona, London, Los Angeles, etc. (Fig. S1). On the other hand, during
Fig. 1. Spatiotemporal variation (panel a and b) and changes in NO2 tropospheric column ($\mu$mol/m$^2$) (panel c) in 20 cities during February 1 to May 11 derived from Sentinel 5 P TROPOMI sensor.
the same period, CO was increased in Cologne and Denver (Fig. S1). Fig. 2 shows the average tropospheric NO$_2$ and CO column values from 1 February to 11 May. The average NO$_2$ (μmol/m$^2$) in 2019 and 2020 was found highest in Tehran, followed by Milan, New York, Paris, Turin, Chicago, Cologne, Philadelphia, etc. The lowest NO$_2$ column values (μmol/m$^2$) are found in Sao Paulo, Brussels, and Denver, respectively. The average CO values (μmol/m$^2$) was found highest in American cities, i.e., New York, Philadelphia, Detroit, Chicago, Los Angeles, while a comparably low tropospheric CO column (μmol/m$^2$) values are seen in Sao Paulo, Denver, Madrid, Barcelona, and Brussels (Fig. 2). Except for a few cities, NO$_2$ and CO column values have been reduced substantially during the study period (Fig. 3, Table 1). For NO$_2$, the highest reduction was detected in Paris (46%), followed by Detroit (40%), Milan (37%), Turin (37%), Frankfurt (36%), Philadelphia (34%),

Fig. 1. (continued).
London (34%), and Madrid (34%), respectively. At the same time, a comparably lower reduction in NO$_2$ is observed in Los Angeles (11%), Sao Paulo (17%), Antwerp (24%), Tehran (25%), and Rotterdam (27%), respectively (Fig. 3). For CO, the maximum reduction was recorded for New York (4.24%), followed by Detroit (4.09%), Sao Paulo (3.88%), Philadelphia (3.45%), Milan (3.17%), Barcelona (2.86%), respectively. At the same time, a positive (increase) changes in CO were observed in Denver (1.92%), Cologne (0.49%), and Rotterdam (0.01%) (Fig. 3). The temporal variability of NO$_2$ is presented in Fig. 4, Fig. 5. Both median and interquartile range (IQR) values in Figs. 4 and 5 suggesting that NO$_2$ was decreased substantially.

Using the in-situ air pollution data, the concentration (μg/m$^3$) of NO$_2$, PM$_{2.5}$, and PM$_{10}$ was evaluated and presented in Fig. 6 and Table 2. NO$_2$ concentration (μg/m$^3$) was found highest in American cities. On the contrary, the concentration of particulate matter (PM$_{2.5}$ and PM$_{10}$) was found highest in European cities. The changes (%) in NO$_2$, PM$_{2.5}$, and PM$_{10}$ concentration from reference values (NO$_2$, PM$_{2.5}$, and PM$_{10}$ concentration from 1 February to 11 May in 2019), was also measured (Fig. 7). The in-situ data suggest that the reduction in NO$_2$ was maximum in Brussels, followed by Paris and London. At the same time, PM$_{2.5}$ and PM$_{10}$ were decreased substantially in London, Frankfurt, and Rotterdam. Additionally, PM$_{2.5}$ and PM$_{10}$ concentration were found to
be increased in Los Angeles. The possible reason for this increment is discussed in the Discussion section.

3.2. Changes in human mobility

Using the Apple human mobility data, the driving, and transit driven mobility was calculated and presented in Fig. 8. Mobility on January 13 was taken as a baseline, and further changes in human mobility during
the study period was calculated from the baseline mobility. The driving counts reduced most substantially in Paris, followed by Madrid, London, Antwerp, and Brussels (Fig. 8, Table S2). Whereas, such changes were comparably low in Chicago, Cologne, Denver, Los Angeles, New York (Fig. 8). Transit counts also reduced considerably in Paris, followed by Utrecht, Sao Paulo, New York, Milan, Chicago, Antwerp, and Brussels (Fig. 8). Google mobility records, which has six mobility components, i.e., retail and recreation, grocery and pharmacy stores, transit, parks and outdoor, workplace visitor, and time spent at home, were also utilized for country-wise assessment of human mobility changes (Fig. S2). Transport related mobilities were reduced most substantially in the Latin American countries, followed by a few Middle East and Southeast Asian countries, and American countries (Fig. S2; Fig. S3). Parks and outdoor activities were found to be reduced maximum in the Latin American countries and South Asian countries (Fig. S3). At the same time, outdoor activities are seen to be increased in a few European countries as well (Fig. S3). The highest reduction in retail and recreation was found in India, Turkey, the UK, and few Latin American countries due to lockdown and associated restrictive measures. (Fig. S4). Considering grocery and pharmacy-related mobilities, the highest reduction is being observed in the Latin American countries and a few European countries. Whereas grocery related mobility was found to be increased in the USA, few African and European countries (Fig. S4). Workplace related mobility is reduced considerably in Peru, Bolivia, India, Spain, Turkey, Saudi Arabia, USA, and Canada (Fig. S5). At the same time, such changes were positive in a few African countries (Mali, Niger, Mozambique, Zambia), Venezuela, and a few island countries (Fig. S5). Finally, using the Google real-time mobility information, another mobility component, i.e., time spent at home, was calculated (Fig. S5). As expected, due to lockdown and mandatory restrictive measures on human activities, people tend to spend more time at home, which also suggests that at most of the countries have taken timely decisions to control the pandemic. Except for a few European countries, peoples around the world limited their outdoor activities, which is supported by the results shown in Fig. S5.

3.3. Lockdown and improving status of air quality

Both public health burden (applied for in-situ data) and externality valuation (applied for Sentinel TROPOMI pollution data) approaches were utilized for assessing economic benefits and economic burden attributed to air pollution led cause-specific mortality (Table 3, Table 4, Table 5, Table S3; Table S4; Table S5). For satellite data-based economic assessment, the default unit \( \mu\text{mol/m}^2 \) was converted to a mass unit (Ton) using the standard mass conversion approach (Borsdorff et al., 2018; Ialongo et al., 2020; Liu et al., 2020). Also, to estimate the total economic benefits of air pollution reduction, the difference in pollution concentration between the current year (1 February to 11 May in 2020) and the preceding year (1 February to 11 May in 2019) was computed. The per-unit economic benefits (US$) due to the reduction of air pollution was found maximum in Sao Paulo (49,709), followed by New York (49,447), Tehran (43,625), London (38,928), Detroit (22,585), Los Angeles (20240), Philadelphia (19188), Madrid (16,413), Chicago (13,222), Milan (10,034), Frankfurt (5854), Turin (5749), Antwerp (5039), Paris (4971), Barcelona (4116), Cologne (3915), Rotterdam (3401), Brussels (1876), and Utrecht (1675) (Table 4). The health burden and associated economic impacts of COVID-19 led reduction in \( \text{NO}_2, \text{PM}_{2.5}, \text{PM}_{10} \) concentrations across selected 20 cities were analyzed and

![Fig. 2. NO2 and CO tropospheric column (μmol/m²) in 20 cities during February 1 to May 11 in 2019 and 2020, derived from Sentinel TROPOMI sensor.](image-url)
Fig. 3. Changes (%) in NO₂ and CO during February 1 to May 11 in 2019 and 2020, derived from Sentinel 5 P TROPOMI data.

Table 1
Total emission (ton) of different air pollutants in 2019 and 2020 derived from Sentinel TROPOMI.

| City       | NO₂  | 2019 | 2020 | Difference (%) | CO    | 2019  | 2020  | Difference (%) |
|------------|------|------|------|----------------|-------|-------|-------|----------------|
| Antwerp    | 1.73 | 1.31 | -24.14 | 215.83        | 212.80 | -1.40 |
| Brussels   | 1.20 | 0.86 | -27.95 | 167.84        | 167.68 | -0.10 |
| Chicago    | 5.55 | 3.88 | -30.09 | 656.87        | 652.04 | -0.74 |
| Cologne   | 3.62 | 2.47 | -31.77 | 426.27        | 428.37 | 0.49  |
| Denver     | 2.97 | 1.98 | -33.43 | 336.58        | 343.06 | 1.92  |
| Detroit    | 3.16 | 1.89 | -40.29 | 409.95        | 393.18 | -4.09 |
| Frankfurt  | 2.14 | 1.36 | -36.36 | 263.32        | 261.39 | -0.73 |
| London     | 12.45| 8.20 | -34.15 | 1644.88       | 1627.06 | -1.08 |
| Los Angeles| 10.63| 9.51 | -10.54 | 1405.58       | 1390.45 | -1.08 |
| Madrid     | 5.18 | 3.42 | -34.03 | 555.52        | 547.84 | -1.38 |
| Milan      | 2.15 | 1.36 | -36.85 | 196.23        | 190.00 | -3.17 |
| New York   | 8.73 | 6.21 | -28.86 | 899.48        | 861.33 | -4.24 |
| Paris      | 1.00 | 0.54 | -45.94 | 112.10        | 109.37 | -2.43 |
| Philadelphia| 3.17| 2.08 | -34.45 | 411.40        | 397.21 | -3.45 |
| Rotterdam  | 2.50 | 1.63 | -26.72 | 342.27        | 342.31 | 0.01  |
| Sao Paulo  | 8.39 | 6.95 | -17.17 | 1139.46       | 1095.22 | -3.88 |
| Tehran     | 25.09| 18.93| -24.54 | 786.14        | 773.67 | -1.59 |
| Turin      | 1.23 | 0.78 | -36.83 | 136.12        | 132.54 | -2.63 |
| Utrecht   | 0.74 | 0.49 | -33.70 | 104.73        | 104.32 | -0.40 |
presented in Table 5. Health impacts are presented in terms of economic burden (indicates the increased levels of air pollution and resulted in cause-specific mortality) and economic benefits (related to reducing air pollution levels and avoided premature deaths). For NO$_2$, economic benefits (Million US$) were found highest in London, followed by New York, Paris, and Chicago. In comparison, monetary benefits were found minimum in Los Angeles and Utrecht. The economic benefits attributed to the reduction of PM$_{2.5}$ and PM$_{10}$ were found highest in London, followed by New York, Paris, Chicago, Frankfurt, respectively. The city-wise population-weighted average concentration (μg/m$^3$), attributed to NO$_2$, PM$_{2.5}$, and PM$_{10}$, were estimated and presented in Fig. 9. The high PWAC value denotes the higher level of long term exposure to NO$_2$, PM$_{2.5}$, and PM$_{10}$ and vice versa. Among the cities, the higher PWAC values (for NO$_2$) were estimated for the US cities (Denver, Detroit, New York, Los Angeles), compared to the European cities considered in this study. For both PM$_{2.5}$ and PM$_{10}$, the higher level of exposure was computed for the European cities (Milan, London, Paris, Antwerp, Barcelona). Due to the reduction of air pollution concentration, the average population exposed to PM$_{2.5}$, PM$_{10}$, and NO$_2$ was reduced substantially for all cities, except few US cities (Los Angeles and Philadelphia). The reduced level of exposure during the study period suggesting a strong synergistic association between controlled human interference and improved air quality across the world. The health burden (HB) estimates suggest that due to the reduction of air pollution, a total of ~1310 (NO$_2$), ~401 (PM$_{2.5}$), and ~430 (PM$_{10}$) premature cause-specific deaths have been averted. The economic benefits of this avoided mortality were measured as ~10, ~3.1, and ~3.3 Billion US$ for NO$_2$, PM$_{2.5}$, and PM$_{10}$.

4. Discussion

4.1. Relevance of satellite remote sensing in air pollution mapping

The ESA Sentinel 5 P TROPOMI real-time pollution data was successfully utilized for evaluating linkages between the temporary cessation of human interferences and improving air quality across the cities.
Sentinel 5 P satellite mission is one of the finest space-borne applications that provide the crucial key information of air quality, Ozone, ultraviolet radiation, and climate monitoring and forecasting (ESA, 2020). TROPOMI widens the application of the satellite air pollution observation and works in line with other global missions, i.e., SCIAMACHY (2002–2012), GOME-2 (since 2007), and OMI (since 2004) (Lorente et al., 2019). This data has been used for many purposes, including air pollution measurement (Borsdorff et al., 2018; Zheng et al., 2019; Shikwambana et al., 2020), epidemiological studies (Chen et al., 2020; Duthell et al., 2020b; Gautam, 2020; Muhammad et al., 2020; Ogen, 2020; Shehzad et al., 2020); monitoring global volcano (Valade et al., 2019), demographic analysis (Kaplan and Yigit, 2020), evaluating sun-induced chlorophyll fluorescence (SIF) (Guantier et al., 2015), estimation of volcanic sulfur dioxide emission (Thyes et al., 2019), etc. In addition, the advent of Google Earth Engine cloud-based functionality in handling the large volume of spatial data facilitates the application of satellite images for timely decision making and offering cost-benefit solutions to many environmental problems. Evaluating the reliability of remote sensing data is always a matter of concern. Many studies across the world have evaluated the reliability of Sentinel 5 P pollution data with ground measurements. Lorente et al. (2019) have examined the reliability of Sentinel TROPOMI tropospheric column NO\textsubscript{2} with in-situ (ground NO\textsubscript{2} boundary layer height over the Eiffel Tower was used in this purpose) data and found a very good agreement ($R^2 = 0.88$) between the two estimates. Griffin et al. (2019) study on validating TROPOMI data with aircraft and surface in situ NO\textsubscript{2} observations over the Canadian oil sands found that the TROPOMI vertical NO\textsubscript{2} column values are strongly correlated ($R^2 = 0.86$) with the aircraft and ground in situ NO\textsubscript{2} observations with a low bias (15–30%).

4.2. Reduced anthropogenic emission and improving air quality status across the cities

Air pollution levels in the urban region are mainly influenced by the local emission of pollutants. For example, Zeng et al. (2018) found that ~75% of the daytime O\textsubscript{3} concentration in Wuhan in summer 2016 was...
caused by localized photochemical formation. In cities, the main air pollutants of concern to public health are particulate matter (PM$_{2.5}$ and PM$_{10}$), NO$_x$, and tropospheric ozone (O$_3$). Both satellite and in-situ data suggest a considerable reduction in air pollution in all the 20 cities considered in this study. However, such reduction is not consistent among the cities: it was found to be very high over the European cities (for all three pollutants, i.e., PM$_{2.5}$ and PM$_{10}$, NO$_x$), and comparably low over the US cities. NO$_x$ concentration was reduced most significantly (>40%) in Brussels and Paris, followed by −35% to −40% in Barcelona and London, −25% to −35% in Rotterdam, Antwerp, Madrid, Utrecht, −15%–25% in Milan, Frankfurt, Detroit, −5%–15% in New York, Denver, Chicago, and less than −5% in Los Angeles, respectively. During the same period, an incremental trend of NO$_x$ was observed in Philadelphia. Several studies also noted that NO$_x$ declined substantially during the COVID-19 time compared to historical years. Berman and Ebisu (2020) observed a statistically significant reduction in NO$_x$ (25.5% reduction with an absolute decrease of 4.8 ppb) in the continental USA from March 13 to April 21 in 2020 compared to average NO$_x$ concentration ($\mu$g/m$^3$) during 2017–2019. Baldasano (2020) estimated that reduction in NO$_x$ concentrations in Barcelona and Madrid (Spain) during the lockdown (March 2020) was 50% and 62%, which is in line with the findings of the present research. Using both satellite and in-situ data, Chen et al. (2020) observed that NO$_x$ concentration had been reduced substantially in China ($12.9 \cdot 9 \mu$g/m$^3$). Venter et al. (2020) have analyzed ground-level measurements from >10,000 air quality stations in 34 countries and recorded a substantial NO$_x$ reduction (60%, 11 $\mu$g/m$^3$ in absolute terms) during the COVID lockdown dates.

In the urban region, NO$_x$ is mainly produced by human activities, including traffic emission, fuel combustion, and partly from industrial emission. EEA (2019) documented that in Europe, transport sector (road transport contributed to 39% of total emission, non-road transport contributed to 8% of total emission) is the most significant contributor to NOx emissions, followed by commercial, institutional and households emission, contributed to 14% of total NO$_x$ emission (EEA, 2019).

Fig. 5. Variation in NO$_2$ tropospheric column (μmol/m$^2$) in the selected cities during the study period in 2019 and 2020, derived from Sentinel 5 P TROPOMI data.
Additionally, in urban regions, the higher level of NO\textsubscript{2} concentration is mostly evident in cities with higher motorized traffic share, industrial regions, and densely populated areas (Zoran et al., 2020). Since we observed a drastic reduction in human mobility, including driving and transit during the observation period, the reduced level of traffic emission could therefore be linked to the decreasing trend of NO\textsubscript{2} observed in the European cities. The reduction of other sectorized emissions, such as industrial emission and commercial, institutional, and household emission, can also be associated with changes in NO\textsubscript{2} observed in these European cities. Kumar et al. (2020a) study on Indian cities noted that among all the influencing factors, including lockdown strictness, switch-off time to halt human activities, local meteorological condition, reduction in road traffic volume was found to be the most influential factor for explaining the variation in air pollution across cities.

Among the cities, the highest NO\textsubscript{2} reduction (41%) was recorded in Brussels. Fierens et al. (2011) reported that air pollution in Brussels is strongly associated with traffic-related pollutants. Apple’s mobility report also shows a substantial reduction in road traffic (47% in driving and 25% in transit) in Brussels. Thus, the reduced level of road traffic volume could explain the noteworthy reduction of NO\textsubscript{2} concentration observed in Brussels. The second highest reduction in NO\textsubscript{2} concentration was recorded in Paris (40.6%). In Paris, traffic and residential sectors are the main source of NO\textsubscript{2} pollution (Connerton et al., 2020). Traffic corresponds to 69% of NO\textsubscript{x} emissions, 36% of PM\textsubscript{10} emissions, and 35% of PM\textsubscript{2.5} emission (Connerton et al., 2020). At the same time, the other contributing factors, such as the residential sector, contributes to 21% of total NO\textsubscript{x} emissions, 41% of total PM\textsubscript{10} emissions, and 49% of total PM\textsubscript{2.5} emissions (Connerton et al., 2020). Other European capital cities, such as London, Barcelona, Madrid have also witnessed a drastic reduction in NO\textsubscript{2} concentration during the studied period. Reduction in road traffic and meteorological factors could be linked with this reduced NO\textsubscript{2} concentration observed in this study (Berman and Keita, 2020; Muhammad et al., 2020; Sharma et al., 2020). The NO\textsubscript{2} concentration in the US cities has not been changed significantly during the study period. Among the six US cities, the negative NO\textsubscript{2} changes were found lowest in Los Angeles (0.16%), followed by
Fig. 6. NO$_2$, PM$_{2.5}$, and PM$_{10}$ concentration (μg/m$^3$) in selected cities during February 1 to May 11 in 2019 and 2020, derived from in-situ data.
traffic volume (Kumar et al., 2020a), uses and mode of domestic energy (Singh et al., 2020), the strictness of the lockdown measures (Singh et al., 2020), including the starting date of lockdown in different cities (Venter et al., 2020), the closure of transportation and industries across the cities. Guevara et al. (2020) observed a similar pattern of air pollution concentration in Los Angeles, Chicago, and Detroit (>25% reduction in Denver, Paris, New York, 5% – 15% – 25% reduction in Paris, Frankfurt, Antwerp, and Philadelphia, 5% – 15% (Detroit, Madrid, Philadelphia). In Milan and Los Angeles, PM_{10} concentration was found to be increased during the studied period. Urrego and Urrego (2020) study examined the effect of lockdown in Paris, Madrid, and Barcelona. Since we utilized both satellite and in-situ data in the assessment, the health impact was presented in terms of ER (excess risk due to exceeding level of air pollution) and health burden (avoided premature mortality due to the reduction of air pollution). The details of RR, ER are given in Table S3, Table S4, Table S5. Combinedly, a total of ~1310 (NO_{2}), ~401 (PM_{2.5}), and ~430 (PM_{10}) cause-specific premature death were averted during the study period, which valued ~16 Billion US$. HB was sharply declined from the previous year’s baseline. This demonstrates the harmonious association between limited anthropogenic appropriations and resulting economic and health (co)benefits. Kumar et al. (2020a) observation in five Indian cities also found a strong positive association between lockdown restrictions and improved health benefits across the cities. However, Kumar et al. also stated that complete/partial closes of business and industries should not be the optimal way of handling air pollution problems; instead, such an estimate can be treated as a mere supposition to reveal the synergistic association between limited human interferences and associated health/economic (co) benefits.

4.3. Human mobility and its association with air pollution

The connection between human mobility and air pollution levels in selected cities were also examined in this research. Results derived from both Google and Apple mobility report suggested that due to the mandatory lockdown and resulted in limited outdoor human activities, mobility has been reduced significantly across the world. This drastic reduction of human mobility could contribute to the reduced level of air pollution observed in the last few months. For most of the cities analyzed, human mobility has been reduced up to 80% from the baseline mobility. The highest reduction in mobility was found in the European cities. To prevent the spread of the pandemic, the authorities in these cities implemented strong preventive measures, which included partial lockdown in different sectors, including restricted outdoor social activities. This mandatory imposition of lockdown has resulted in a reduced

| Table 2 |
| --- |
| NO_{2} concentration (µg/m³) during the study period (1 February to 11 May) in 2019 and. |
| 2019 | 2020 | Δ NO_{2} (%) | 2019 | 2020 | Δ PM_{10} (%) | 2019 | 2020 | Δ PM_{2.5} (%) |
| Antwerp | 33.83 | 22.76 | −32.72 | 31.76 | 25.40 | −20.02 | 18.96 | 12.81 | −32.44 |
| Barcelona | 35.77 | 22.71 | −36.50 | 24.23 | 19.74 | −18.52 | 13.55 | 9.47 | −30.10 |
| Brussels | 31.04 | 18.32 | −40.98 | 20.73 | 15.69 | −24.28 | 14.48 | 9.36 | −35.32 |
| Chicago | 52.35 | 46.56 | −11.08 | 25.73 | 20.99 | −18.42 | 9.21 | 8.56 | −7.03 |
| Denver | 60.18 | 53.41 | −11.26 | 27.48 | 20.91 | −23.92 | 8.44 | 6.35 | −24.81 |
| Detroit | 50.68 | 40.54 | −20.01 | 17.93 | 16.06 | −10.44 | 9.18 | 8.15 | −11.22 |
| Frankfurt | 28.37 | 22.14 | −21.95 | 20.84 | 14.86 | −28.71 | 13.55 | 9.47 | −30.10 |
| London | 41.83 | 25.39 | −39.31 | 25.40 | 16.33 | −35.73 | 15.58 | 9.88 | −36.56 |
| Los Angeles | 48.81 | 48.73 | −0.16 | 18.55 | 21.28 | 14.70 | 7.42 | 8.46 | 13.98 |
| Madrid | 35.69 | 24.16 | −32.32 | 16.78 | 14.95 | −10.88 | 9.00 | 8.69 | −3.40 |
| Milan | 44.14 | 34.45 | −21.96 | 31.44 | 31.79 | −1.13 | 21.21 | 22.61 | 6.57 |
| New York | 50.98 | 43.94 | −13.80 | 23.17 | 22.78 | −1.39 | 13.61 | 8.71 | −36.01 |
| Paris | 44.16 | 26.25 | −40.55 | 25.03 | 18.17 | −27.40 | 16.87 | 13.07 | −22.55 |
| Philadelphia | 42.35 | 43.67 | 3.11 | 17.34 | 14.84 | −14.40 | 7.99 | 7.76 | −2.96 |
| Rotterdam | 31.96 | 21.20 | −33.66 | 23.17 | 23.66 | −2.28 | 13.61 | 8.71 | −36.01 |
| Utrecht | 24.29 | 17.06 | −29.77 | 21.18 | 17.11 | −19.21 | 13.98 | 9.79 | −29.95 |
Fig. 7. Changes in NO2, PM2.5, and PM10 concentration (%) during February 1 to May 11 in 2019 and 2020, derived from in-situ data.
Fig. 8. Changes in mobility due to lockdown and resulted in restriction in the selected cities.
with odd digit (1, 3, 5, 7, 9) registration numbers will be allowed on introducing an roads on odd dates and vehicles with even digit (0, 2, 4, 6, 8) registration highly polluted cities if no other alternatives are feasible to adopt. A study had observed a 40% level of traffic volume in cities (Fig. 8, Table S2). The periodic and temporary lockdown can also be adopted in within the first week of lockdown. He et al. (2020) study on short-term policies, air pollution in the urban regions can be minimized to a certain thus suggests that by introducing sustainable transport plans and pol

level of traffic volume in cities (Fig. 8, Table S2). The mobility analysis thus suggests that by introducing sustainable transport plans and policies, air pollution in the urban regions can be minimized to a certain extent. The periodic and temporary lockdown can also be adopted in highly polluted cities if no other alternatives are feasible to adopt. A similar strategy has already been adopted by New Delhi Government by introducing an “odd/even” transport scheme where private vehicles with odd digit (1, 3, 5, 7, 9) registration numbers will be allowed on roads on odd dates and vehicles with even digit (0, 2, 4, 6, 8) registration numbers can use the vehicles on even dates. Additionally, Mahato et al. study had observed a 40%-50% improvement in air quality in Delhi within the first week of lockdown. He et al. (2020) study on short-term impacts of COVID-19 lockdown on urban air pollution has found that within a week, the AQI in the locked-down cities in China has been reduced by 19.84 points (PM$_{2.5}$ goes down by 14.07 μg/m$^3$) compared to the cities where lockdown has not been implemented strictly. The findings suggest an increased clean air ecosystem services in cities under the cessation of human activities.

5. Conclusion

We made an effort to investigate the positive effects of COVID-19 lockdown restrictions on the reduction of NO$_2$, PM$_{2.5}$, and PM$_{10}$ concentration. Different valuation methods, including median externality and public health burden, were incorporated into the economic valuation to assess the health impact and economic benefits of avoided mortalities. Both satellite and ground-based estimates are exhibiting an affirmative association between controlled human interference and improved air quality. The outcome of this research demonstrates the strong connection between the decline in traffic volume and reduction of NO$_2$, PM$_{2.5}$, and PM$_{10}$ emission across the cities. This also suggests that the controlled motorized traffic pollution and limiting other unsustainable human activities could be the most effective ways of improving the air quality status of a city. Though the selected pollutants have shown a substantial reduction in all the 20 cities analyzed, there has been an irregularity in the reduction of air pollutants found. Many factors, including meteorological factors, start time of lockdown restrictions, the strictness of lockdown measures, volume of road traffic, other point sources of localized emission, could be linked to this varying concentration and reduction of air pollution observed in the selected cities. The outcome of this study can be a reference to introduce new public policies for promoting adaptive socio-ecological models to understand the synergies and trade-offs between the reduced human interventions and the environmental health of cities systematically. Further research in this direction is needed to explore this synergistic association more explicitly.

Credit author statement

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### Table 3

| Pollutants | Min | Median | Mean | Max |
|------------|-----|--------|------|-----|
| CO         | 2   | 956    | 956  | 1931|
| NO$_x$     | 404 | 1949   | 5149 | 17,468|
| PM$_{10}$  | 1747| 5149   | 7907 | 29,788|

### Table 4

| City          | NO$_x$  | CO  | Overall Economic Benefit | Economic Benefit | Economic Benefit |
|---------------|---------|-----|--------------------------|------------------|------------------|
|               |         |     | (US$)                    | (US$)            | (US$)            |
| Antwerp       | 2145    | 2894| 3039                     |                  |                  |
| Barcelona     | 1251    | 2866| 4116                     |                  |                  |
| Brussels      | 1720    | 156 | 1876                     |                  |                  |
| Chicago       | 8606    | 4616| 13,222                   |                  |                  |
| Cologne       | 5924    | –   | 3915                     |                  |                  |
| Denver        | 5115    | –   | 1075                     |                  |                  |
| Detroit       | 6550    | 16,036| 22,585                  |                  |                  |
| Frankfurt     | 4007    | 1847| 5854                     |                  |                  |
| London        | 21,888  | 17,040| 38,928                  |                  |                  |
| Los Angeles   | 5771    | 14,469| 20,240                  |                  |                  |
| Madrid        | 9073    | 7340| 16,413                   |                  |                  |
| Milan         | 4083    | 5951| 10,034                   |                  |                  |
| New York      | 12,976  | 36,471| 49,447                  |                  |                  |
| Paris         | 2362    | 2608| 4971                     |                  |                  |
| Philadelphia  | 5625    | 13,563| 19,188                  |                  |                  |
| Rotterdam     | 3437    | –   | 3401                     |                  |                  |
| Sao Paulo     | 7415    | 42,294| 49,709                  |                  |                  |
| Tehran        | 31,702  | 11,923| 43,625                  |                  |                  |
| Turin         | 2328    | 3421| 5749                     |                  |                  |
| Utrecht       | 1279    | 396 | 1675                     |                  |                  |

### Table 5

| City          | NO$_x$  | PM$_{2.5}$ | Economic Benefit | PM$_{2.5}$ | Economic Benefit | PM$_{10}$ | Economic Benefit |
|---------------|---------|------------|------------------|------------|------------------|-----------|------------------|
|               |         |           |                  |            |                  |           |                  |
| Antwerp       | 616     | 413        | 179              |            |                  |           |                  |
| Barcelona     | 1381    | 167        | 56               |            |                  |           |                  |
| Brussels      | 296     | 167        | 56               |            |                  |           |                  |
| Chicago       | 5486    | 1326       | 89               |            |                  |           |                  |
| Denver        | 1475    | 295        | 70               |            |                  |           |                  |
| Detroit       | 1339    | 331        | 35               |            |                  |           |                  |
| Frankfurt     | 1107    | 636        | 181              |            |                  |           |                  |
| London        | 12,815  | 6534       | 2078             |            |                  |           |                  |
| Los Angeles   | 8098    | 1682       | 224              |            |                  |           |                  |
| Madrid        | 2515    | 810        | 26               |            |                  |           |                  |
| Milan         | 1720    | 1023       | –59              |            |                  |           |                  |
| New York      | 16,907  | 3406       | 502              |            |                  |           |                  |
| Paris         | 2742    | 1329       | 277              |            |                  |           |                  |
| Philadelphia  | 2778    | 694        | 20               |            |                  |           |                  |
| Rotterdam     | 337     | 175        | 60               |            |                  |           |                  |
| Utrecht       | 318     | 216        | 61               |            |                  |           |                  |
Declaration of competing interest

The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

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Appendix A. Supplementary data

Supplementary data to this article can be found online at https://doi.org/10.1016/j.envres.2021.110927.

Fig. 9. Population Weighted Average Concentration (PWAC) (µg/m³) for different pollutants in the selected cities.

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