Environmental, evolutionary, and ecological drivers of slow growth in deep-sea demersal teleosts

Jesse A. Black1,*, Anna B. Neuheimer1,2, Peter L. Horn3, Di M. Tracey3, Jeffrey C. Drazen1

1Department of Oceanography, School of Ocean Earth Science and Technology, University of Hawaii, Honolulu 96822, USA
2Aarhus Institute of Advanced Studies & Department of Biology, Aarhus University, Aarhus 8000, Denmark
3National Institute of Water & Atmospheric Research Ltd (NIWA), Wellington 6241, New Zealand

ABSTRACT: The deep sea (>500 m ocean depth) is the largest global habitat, characterized by cool temperatures, low ambient light, and food-poor conditions relative to shallower waters. Deep-sea teleosts generally grow more slowly than those inhabiting shallow water. However, this is a generalization, and even amongst deep-sea teleosts, there is a broad continuum of growth rates. The importance of potential drivers of growth rate variability amongst deep-sea species, such as temperature, food availability, oxygen concentration, metabolic rate, and phylogeny, have yet to be fully evaluated. We present a meta-analysis in which age and size data were collected for 53 species of teleosts whose collective depth ranges span from surface waters to 4000 m. We calculated growth metrics using both calendar and thermal age, and compared them with environmental, ecological, and phylogenetic variables. Temperature alone explained up to 30% of variation in the von Bertalanffy growth coefficient ($K_{yr^{-1}}$), and 21% of the variation in the average annual increase in mass (AIM, %), a metric of growth prior to maturity. After correcting for temperature effects, depth was still a significant driver of growth, explaining up to 20 and 10% of the remaining variation in $K$ and AIM, respectively. Oxygen concentration also explained ~11% of remaining variation in AIM following temperature correction. Relatively minor amounts of variation may be explained by food availability, phylogeny, and the locomotory mode of the teleosts. We also found strong correlation between growth and metabolic rate, which may be an underlying driver also related to temperature, depth, and other factors, or the 2 parameters may simply covary as a result of being linked by evolutionary pressures. Evaluating the influence of ecological and/or environmental drivers of growth is a vital step in understanding both the evolution of life history parameters across the depth continuum as well as their implications for species’ resilience to increasing anthropogenic stressors.

KEY WORDS: Fish · Teleost · Growth · Pace of life · Evolution · Life history · Metabolism · Temperature · Deep sea · Demersal · Marine
1. INTRODUCTION

Deep-sea teleosts (referred to hereafter as fishes) generally have high ages-at-maturity, slower growth rates, lower metabolic rates, and greater longevity than shallow-water fishes (Koslow 1996, Caillet et al. 2001, Drazen & Haedrich 2012). Despite these generalizations, life history parameters vary considerably across the depth continuum. For example, the Patagonian toothfish Dissostichus eleginoides (median depth of ~1100 m) matures between 5 and 7 yr of age (Everson & Murray 1999, Horn 2002), while the black cardinalfish Epigonus telescopus (median depth ~800 m) matures around 36 yr (Tracey et al. 2000, 2017, Dunn 2009). Differences in longevity, maturity, and growth rate may reflect the cumulative influence of phylogeny, evolutionary pressures such as predation, and a variety of environmental drivers such as pressure, temperature, oxygen, and food availability, all of which change with depth in most ocean habitats.

A variety of environmental drivers, temperature and food supply in particular, may affect growth rates of shallow-water fishes. Temperature has been shown to be a primary driver. The growing degree-day (GDD °C × day), i.e. the product of habitat temperature and calendar age of fish in days, has successfully explained differences in growth rates across populations of Atlantic cod Gadus morhua and other fishes (Braten & Guy 2002, Neuheimer & Taggert 2007, 2010, Venturelli et al. 2010, Shackell et al. 2019). In much of the deep sea, temperature and growth rates decline rapidly with depth, and so temperature has been frequently hypothesized to drive the observed patterns (Drazen & Haedrich 2012, McClain et al. 2012). Flux of particulate organic carbon (POC), the primary basis of the food web in the deep ocean, also declines exponentially with depth (e.g. Martin et al. 1987, Lutz 2007). Thus, the declining availability of food with depth may also play a role in growth (e.g. Rigby & Simpfendorfer 2015), as accumulating mass requires energy intake and, at least in aquaculture settings, feeding level has a large influence on growth rates (reviewed in Persson & De Roos 2006). In fishes’ natural environment, however, ecosystem-level changes in food supply may primarily constrain population size before affecting interspecific growth rate; as available food increases, the carrying capacity of a given population increases, while the amount of energy available to individuals for growth and other biological processes may remain relatively constant, depending on relationships between average body size and foraging efficiency (Sebens 1987, McClain et al. 2012).

Metabolic rate has been found to correlate strongly with growth rates. Metabolic rate, which scales to body mass and temperature, represents the rate of biochemical reactions in the organism (Brown et al. 2004, McClain et al. 2012), including the distribution of biochemical resources into the processes of growth, reproduction, and survival. It is unclear whether metabolic rate mechanistically constrains maximum growth rate (the ‘performance model’), or rather represents the fraction of energy devoted to maintenance costs, with the remaining energy diverted to growth and activity (the ‘allocation model’), as studies on teleosts have found support for both models (reviewed in Careau et al. 2008). If metabolic rate does positively and causally correlate to growth, then any environmental conditions affecting metabolism would therefore indirectly affect growth rate. Temperature clearly declines with depth in most ocean basins which could then reduce metabolic rates with concurrent reductions in growth rate. Oxygen is a key reactant in aerobic metabolism, and if growth rate is indeed constrained by or linked to metabolic rate, it would follow that aerobic metabolic reactions that facilitate generation of new biomass can only operate as quickly as oxygen can be made available. However, metabolic rate has been shown to be more strongly associated with gill surface area than the ambient concentration of oxygen itself, and thus rapid metabolism and growth in oxygen-poor waters may simply be a matter of possessing specific adaptations for efficient oxygen extraction and delivery (Friedman et al. 2012). Finally, the visual interactions hypothesis suggests that after accounting for temperature and body size effects, the decline in metabolism with depth in fishes (and several midwater groups) to a depth of about 1000 m reflects a decrease in selective pressure for rapid locomotory capacity as visual predation becomes less common with decreasing light levels (Childress 1995). In short, temperature, oxygen, and light levels may all affect metabolic rate, depending largely on the depth in question, and may therefore be indirect but important drivers of growth.

Many of these variables change with depth, often covarying with one another and/or varying between regions of the ocean (Paulmier & Ruiz-Pino 2009). Therefore, distinguishing the importance of one factor over another will require knowledge of life histories from multiple regions of the oceans to untangle as much covariation as possible. Studies of age and growth require extensive sampling effort, time, and labor, and these logistical demands increase sharply as deeper dwelling organisms are targeted. The progress of age and growth studies for many deep-sea
species may therefore be outpaced by anthropogenic impacts, and as such, a predictive framework of growth is required in the short term. To address these factors, we compared life history parameters from studies on populations of deep-sea demersal fishes and some phylogenetically related shallow-water species to environmental variables from each species’ habitat to establish a framework for estimating growth patterns and life histories. Similar analyses have been performed on octopods, freshwater fishes, and sharks (Rigby & Simpfendorfer 2015, Schwarz et al. 2018, de Santana et al. 2020) to explain the effect of temperature and other ecological and environmental variables on life history parameters, but to our knowledge, this is the first meta-analysis to incorporate environmental conditions in explaining growth of marine demersal fishes across depth.

It may be that variation in growth across demersal fishes is not directly explained by environmental conditions. Rather, it may relate to the relatively higher fitness of a ‘slow’ life history strategy with increasing depth, with the deep sea favoring slow metabolism, low growth rates, and long maturation times relative to the shallows. A spectrum of ‘fast vs. slow’ life history strategies has been observed across a diversity of animals, with ‘faster’ organisms tending to have a suite of traits including higher growth rates, faster metabolisms, shorter maturation times, increased mortality, bolder behavior, etc. relative to ‘slower’ organisms (Mittelbach et al. 2014, Auer et al. 2018, Damsgård et al. 2019; reviewed in Réale et al. 2010). Much of this research has focused on intraspecific patterns, but it may apply to interspecific comparisons too. We discuss the implications of evolutionary drivers even though their quantification is much more challenging than environmental determinants in the deep sea or any environment.

2. METHODS

2.1. Growth data and metrics

Data on fish growth and life history were extracted from a variety of sources (see Table 1). Forty published studies and 2 national oceanographic institutions (Northwest Fisheries Science Center [NWFSC; accessed via https://www.nwfs.noaa.gov/data/map] operating out of the USA, and the National Institute of Water and Atmospheric Research [NIWA] out of New Zealand) also provided data. Raw data were used when possible. When raw data were unavailable, data were digitized from plots of age vs. size, or from tables containing mean size-at-age. When available, other relevant spatial and environmental data were also collected, including depth, latitude and longitude, and temperature and/or oxygen concentration measured on the fishing gear. We attempted to minimize the effects of covariation between depth and temperature by collecting data from a global set of locations, including the Mediterranean Sea, where temperature remains relatively constant (~13°C) across depth. Only teleost fishes were included in this analysis. Chondrichthians were omitted as there is a scarcity of age data for this group and they have fundamentally different life histories and physiologies (e.g. Thiberg & Speers-Roesch 2016).

All age data used in this meta-analysis originated from counts of otolith annuli. To be accepted for use in our meta-analysis, otoliths must have been at least polished/ground/broken to reveal the central nucleus before age estimation. Age data from whole otoliths has been widely shown to severely underestimate actual fish age for multiple species, particularly those that are long-lived (reviewed in Campana 2001). For this reason, we did not include age data for mature fishes derived from whole otolith age counts, with the only exception being Beryx splendens. Studies have validated the use of whole otoliths as an accurate aging method for B. splendens via marginal frequency analysis and direct comparison of annuli counts from whole and sectioned otoliths (Massey & Horn 1990, Lehodey & Grandperrin 1996). Some studies in this analysis used whole otoliths to estimate ages of young fish but switched to other methods (break and burn or transverse sectioning) for fish above a particular size. Only studies with age data for at least 90 individual fish were used in this analysis. In some cases, data for more than one population of the same species was collected, and growth parameters and environmental conditions were estimated separately for these distinct populations. The only exception is Coryphaenoides acrolepis. For this species, in order to obtain enough observations to fit growth parameters to age and size data, data were combined from Andrews et al. (1999) and Matsui et al. (1990), which were collected primarily from the California slope with some data from off Oregon and Washington. The catch locations for 7 species from the NWFSC Groundfish Survey, which spans the entire west coast of the USA, were initially split into North and South sections to test for differences in environmental parameters and growth rates between the 2 regions. All these species are mobile, and no evidence for multiple, isolated populations of these fishes along the western USA was found in the literature. Of these
7 species, only Merluccius productus showed significant differences in growth metrics (~30% difference) between fish from north and south of 42.5° N, so these groups were parameterized as different populations (see Table 1). The rest of the species from the NWFSC Groundfish Survey were parameterized as one contiguous population because no significant differences in growth parameters were found.

Growth rate in fish has been traditionally estimated using the von Bertalanffy (1938) growth curve: \( L_t = L_\infty \left( 1 - e^{-K(t-t_0)} \right) \), where \( L_t \) is length of the fish at time \( t \), the \( x \)-intercept \( t_0 \) is the theoretical age at a length of zero, and \( K \) represents the exponential coefficient of growth until a maximum asymptotic length \( (L_\infty) \). A larger value of \( K \) corresponds to a more rapid increase in length to the maximum length of the fish. These parameters were calculated for each population. Bootstrapped 95% confidence intervals were calculated for each parameter. For data sets with few young fish, this method can produce biologically improbable curves (e.g. a population with predicted lengths of ~20 cm at age 0). For this meta-analysis, any von Bertalanffy curve with a \( t_0 \) with absolute value >2.5 was recalculated with a \( t_0 \) fixed at −0.5. Fixing the \( t_0 \) parameter at −0.1 or −0.5 has been employed in growth studies on species with a relative scarcity of juvenile data in order to ensure predicted fish length is positive and nonzero at hatching, producing more biologically realistic values of \( K \) (Horn et al. 2012, Horn & Sutton 2015). Here, we chose −0.5 as the fixed value, though fixing at −0.1 did not yield significantly different \( K \) values. Still, parameter values should be observed with some caution (Lorenzen 1996). See Supplement 1 for growth curves and Supplement 2 for parameter values and confidence intervals with and without fixed values of \( t_0 \) (all Supplements for this article are at www.int-res.com/articles/suppl/m658p001_supp/).

Growth was also calculated as an average increase in mass (AIM; as %) per year during the prematurity phase, to provide an estimate of the maximum rates at which each species can accumulate mass prior to experiencing the energy expenditures associated with reproduction. When available, age at 50% maturity was collected from the literature (Supplement 2). When age at 50% maturity was not available, we used the median between minimum and maximum age at maturity or estimated a likely age at maturity from published lengths at maturity using von Bertalanffy growth curves (see Table 1). If the age at maturity differed between sexes, the younger age was used as the cut-off for the end of the prematurity phase. When mass data were not available, predicted mass was calculated from fish length using published length–weight conversions (Supplement 2). For each species, the relationship between age and predicted or observed mass was modeled using a generalized additive model (GAM) to predict mass-at-age during the prematurity phase of growth (error distribution: Gamma). These GAM fits were assessed for their ability to give rise to the data by comparing residuals between data and simulation estimates from 1000 simulations to check for uniformity or outlier errors using the R package ‘DHARMa’ (Hartig 2019). To avoid overfitting, the complexity of the curve shape was increased incrementally until simulated fits did not produce these errors. A more detailed description of the GAM fitting process is in the ‘Supplementary Methods’ (Supplement 3). Curves fit for the Atlantic populations of the abyssal Coryphaenoides armatus, the deepest species in our data set, did not meet our thresholds for any fitting approach due to a small data set. However, data for this population was retained to investigate the influence of this abyssal fish on patterns in growth. Results for AIM are reported with and without inclusion of C. armatus. To capture only the phase of rapid growth during prematurity, any decline in predicted mass observed as the fish approached maturity was trimmed. The remaining values of predicted mass-at-age were then used to calculate the AIM.

Depth ranges were summarized from the literature for each species (Supplement 2). Whenever possible, we used published depth ranges that were estimated in the same general area as the catch locations for studies used for this meta-analysis. Several of the species included in this meta-analysis have improbably wide absolute minimum and maximum depths of occurrence. For example, Antimora rostrata has been observed/collected at depths from 350–3000 m, but generally lives at 1200–2500 m (Cohen et al. 1990, Froese & Pauly 2019). In order to constrain predictions of temperature, oxygen, and other environmental parameters to conditions likely experienced by most members of each species, we used these narrower, ‘usual’ depth ranges of high abundance for each species in our analyses rather than the absolute observed depth range (as in Drazen & Haedrich 2012). When modal histograms of abundance vs. depth were available in the literature, the minimum and maximum depth range was set to where abundance was 25% of the maximum abundance. This procedure was also used if histograms of presence in percent of tows vs. depth were available (i.e. Anderson et al. 1998). The median value between minimum and maximum usual depths was also used as a predictor. Thus, for each population, this analysis included 3 depth metrics: minimum, median, and maximum usual depth.
As temperature has been shown to explain a great deal of variation in growth rate in shallow-water species, it was predicted to be a primary driver in both growth metrics, \( K \) and AIM. To decouple the effect of temperature from other potential drivers that also closely correlate with increasing depth, temperature-adjusted \( K \) and AIM were created by calculating versions of each growth metric using growing degree-years (GDY, i.e. the integrated thermal age), instead of calendar age. Here we use GDY, the product of fish age in years and predicted temperature of habitat, to obtain a metric of thermal age for each fish. Thermal ages were then used to create temperature-adjusted versions of \( K \) and AIM, denoted here as \( K_{\text{GDY}} \) and \( \text{AIM}_{\text{GDY}} \). These 2 metrics were analyzed alongside \( K \) and AIM in statistical analyses with environmental and ecological variables.

Temperature at the location of fish captures was not available in most studies. An average temperature experienced by the fish throughout their lifespan was predicted using data from World Ocean Atlas (WOA) (1° horizontal resolution with 5 m vertical resolution near the surface down to 100 m resolution at abyssal depths, decadal averages from 1955–2012; Locarnini et al. 2013). Temperature observations across each species’ usual depth range at the location of catch were averaged. Species that live in habitats that contradict the inverse relationship of depth and temperature in the world ocean were classified as follows for comparison: cold, shallow-water fishes (≤300 m maximum usual depth and <4°C average temperature of habitat) and deep, warm-water fishes (≥300 m minimum usual depth and >10°C average temperature of habitat).

Oxygen concentration (ml l\(^{-1}\)) data were gathered from WOA (1° horizontal resolution with 5 m vertical resolution near the surface down to 100 m resolution at abyssal depths, decadal averages from 1955–2012; García et al. 2013). Oxygen concentration observations across each species’ usual depth range at the capture locations were averaged.

Estimating food availability is a great challenge and depends upon species-specific feeding preferences and foraging strategies (Drazen & Sutton 2017) as well as prey population dynamics. This information was lacking for nearly all species in this study. Therefore, we used POC flux to the seafloor estimated by the Lutz model (Lutz et al. 2007) as a proxy for food availability. POC flux has been used as a proxy for food supply in similar, global meta-analyses (Woolley et al. 2016). POC flux is the basal energy source for most seafloor habitats (except chemosynthetic ecosystems), and therefore supports the production of biomass in benthic habitats, which is theoretically available for fishes to feed upon. The procedure for estimating POC flux to the seafloor for each individual species’ catch location was adapted from methods described by Lutz et al. (2007). Net primary production (NPP) was estimated using global Standard Vertically Generalized Production Model (VGPM) data from Oregon State University’s Ocean Productivity portal (www.science.oregonstate.edu/ocean.productivity/custom.php) at a resolution of 1/6th of a degree. Bathymetry was extracted from the Navy ETOP0 data base (ETOPO5 5 x 5 minute bathymetry; http://iridl.ldeo.columbia.edu/SOURCES/.WORLDBATH/.bathfiles.html), and any observations of chlorophyll along the 30 m isobath or shallower were omitted to avoid unnecessary backscatter, as done in previous published estimations of primary production (Gove et al. 2016). NPP at 2 yr intervals across the years 2007–2017 was averaged across the catch location bounds for each population, resulting in a decadal average value of NPP for each population. Seasonal Variation Index was calculated as described in Lutz et al. (2007). A global average euphotic zone depth of 110 m was used (Lima et al. 2013). Three separate Lutz POC fluxes were predicted for the minimum, median, and maximum depths for each population within their catch boundaries.

Lutz POC flux is dependent on measurements from sediment traps, which have been found to underestimate the inputs of organic carbon to the seafloor, particularly on continental slopes (Smith 1992, Smith & Demopoulos 2003). For this reason, we also gathered sediment community oxygen consumption (SCOC) rate data, a measure of remineralization rates of organic material, which has also been used to estimate ecosystem food availability (Jahne 1996, Stratmann et al. 2019). We used the SCOC database collated by Stratmann et al. (2019) to collect SCOC rates within the bounds of each species’ catch location and averaged them at 50 m intervals across the depth range of each species. These 50 m binned averages were then averaged to provide a mean SCOC value representative of each entire habitat. As these data are sparse geographically and rely on individual measurements rather than extrapolation from satellite color data, representative lifespan SCOC rates could only be calculated for 21 species.

For a subset of fishes, published data from studies on fish metabolic rate were available for comparison to growth metrics. Specifically, published activities of citrate synthase (CS) enzyme activity for various species were collected (Sullivan & Somero 1980, Siebenaller et al. 1982, Drazen et al. 2015, Saavedra et al.
2.2. Ontogenetic migrators

Ten species in this meta-analysis are known to ontogenetically migrate deeper with age/size (see Table 1), complicating estimation of environmental predictors. For these fishes, estimation of environmental predictors was modified to account for changes in depth across the fish’s lifespan. Environmental predictors were estimated individually, based on the unique depth of each individual fish given its capture depth (or mean size-at-age data point, when only mean data were available). When individual depth-of-capture was not available, published regressions were used to estimate likely depth at a given size/age. For *S. alascanus*, published probabilities of depth by length (Jacobson et al. 2001) were used to model predicted length vs. depth (GAM; family: gamma; link: log). When detailed depth information was available (primarily for data from NWSC), temperature was estimated for each ontogenetic migrator by collecting WOA temperature data within each study location and using a model (GAM; family: gamma; link: log) to predict temperature based on each ontogenetic migrator’s predicted or observed depth and latitude/longitude of catch. This process was repeated to estimate oxygen concentration for each migrator. Lutz POC flux was calculated at the predicted or observed depth of each individual ontogenetic migrator (or mean size-at-age datapoint when only mean data were available).

Thus, for each ontogenetically migrating species, mean conditions were available for each year of the species as it moved downslope.

The descriptors ‘prematurity’ and ‘lifespan’ are used next to environmental predictors to specify whether the underlying data for ontogenetic migrators reflects their early, relatively shallow-water period of life, or their entire lifespan, respectively. For ontogenetic migrators, these predictors are each effectively weighted by time (years) spent at particular depths/conditions. For non-ontogenetic migrators, these 2 versions are identical, as these fish are expected to be distributed relatively uniformly in age/size across their depth range. Collected SCOC rates within the catch locations and depth ranges of ontogenetic migrators were not numerous enough to create regressions of SCOC vs. depth and predict SCOC at the depth of each individual ontogenetic migrator. To resolve this issue, we averaged SCOC at the prematurity depth range of each ontogenetic migrator (average individual depth of capture or predicted depth during prematurity, ±SD of the depth of capture). Data were only sufficient to calculate these prematurity SCOC rates for ontogenetic migrators living along the US west coast (*S. alascanus* and species collected in the NWSC Groundfish Trawl Survey; see Table 1). For metabolic rate comparisons, 2 versions of temperature-adjusted CS activities were created by adjusting CS10 values to the temperature experienced by each species across its lifespan (‘lifespan CS10’), and CS activity rate adjusted to the prematurity temperature experienced by each species (‘prematurity CS10’; see ‘Adjusted metabolic rates to temperature and mass of habitat’ in ‘Supplementary Methods’ in Supplement 3 for the equations used).

In terms of calculating thermal age for ontogenetic migrators, GDY was calculated to represent shifting temperature regimes across their lifespan by fitting a generalized linear model (GLM) of age vs. predicted or observed temperature (family: gamma; link function: log). In multiple species, males and females were observed to have statistically different (p <
0.05) age vs. temperature curves, suggesting different rates of downslope migration between sexes. For this reason, separate GLMs were fit to males and females for most ontogenetic migrators. To calculate the cumulative GDY for a fish species, the curve produced by the respective GLM equation for the fish’s species and/or sex was integrated with bounds beginning from age 0 up to the individual fish age. This integration produces the area under the age vs. temperature curve, and thus calculates a thermal age for the individual fish that accounts for the shifting thermal regime encountered by these ontogenetic migrators. See Fig. S1 in Supplement 3 for an example of how GDY was calculated for ontogenetic migrators and non-migrants.

2.3. Statistical analysis

All growth metrics were compared with all environmental predictors in 2 ways: pairwise GLMs (with gamma error distribution), each containing only one growth metric as response and one environmental variable, and with multivariate generalized linear mixed models (GLMMs; gamma error distribution), each containing one growth metric as response, multiple environmental variables, and a random effect for phylogenetic family. The former method was used primarily to identify which metric of each predictor (i.e. maximum depth vs. median depth, prematurity temperature vs. lifespan temperature, etc.) best explained growth and informed which should be included in subsequent GLMM analysis. As such, the results of these pairwise comparisons are not discussed at length here. The latter approach was used to determine which combinations of predictors (including the null model) best explained the growth rate data. In these analyses, a gamma error distribution was chosen because the long-term growth rates of these animals must be greater than zero, and this error distribution does not allow residuals to be negative. Due to paucity of SCOC and metabolic rate data, these predictors were only analyzed in pairwise regressions, and were not included in the full GLMM analysis. However, to see how much variation could potentially be explained by including metabolic rate data, the top model for explaining each growth rate was rerun including prematurity CS_H, the metabolic rate predictor with the strongest relationships with growth rate in pairwise analysis, as a predictor. Environmental predictors estimated from both the prematurity phase of life and across the entire lifespan were used in the pairwise regressions, as well as in GLMM analysis for K and K_GDY. Only prematurity predictors were used in multivariate analysis for AIM and AIM_GDY, as these growth metrics reflect only the prematurity phase of growth. For numerical predictors, either a log link function or inverse link function was used depending on which form of the GLM created residuals that were more evenly distributed and without clear patterns. When a clear difference in residual patterns across link functions was not observed, the model favored by Akaike’s information criterion with correction for small sample size (AICc) was chosen. AICc is a scoring method used to rank models without favoring those that are overly complex, as the AICc is decreased with increasing explanatory power but is increased by model complexity. Models with AICc scores within 2 of one another are assumed to be equally likely to be the ‘true’ model. AICc was used in these analyses in the model selection process to maintain balance between over-fitting and under-fitting the data.

Mixed model analysis and subsequent model selection was repeated for each growth metric with and without inclusion of the abyssal grenadier C. armatus, as this is the only abyssal fish in the data set and is thus a valuable point of comparison to other fishes, but is also a notable outlier in AIM and AIM_GDY. For the mixed models where there was more than one version of a predictor (i.e. minimum usual depth vs. median usual depth vs. maximum usual depth), the one that explained the most variation for each growth metric in their respective pairwise regression (see Table 2) was used.

To reduce the influence of covariation among predictors in mixed models, predictors were stepwise removed from each GLMM until the variation inflation factor (VIF) of each predictor was below 3. VIF thresholds of lower than 10 have been suggested, with more conservative cutoffs in high-collinearity data sets set at 2 or 3 (Zuur et al. 2010). The remaining GLMs were each run through the R function ‘dredge’ (package ‘MuMIn’; Barton 2019), which creates an individual GLMM for every possible permutation of predictors in the global model and returns a table with the models ranked by AICc.

3. RESULTS

In total, 74 teleost growth studies or data sources were considered for this meta-analysis. Of these, 21 were eliminated due to use of only whole otoliths and/or scales in the aging process. Another 12 were eliminated for quality control reasons outlined in the Sup-
plemory Methods (Supplement 3). Thus, data were included from 42 sources: 40 published studies and 2 national oceanographic institutions (NIWA and NWFSC), representing 53 species over 7 orders of teleost fishes (Table 1). Among these species, 22 are shallow-water fishes (maximum usual depth <500 m), with most others with median depths of 500−3150 m, and a maximum depth of 4000 m (Fig. 1). Temperatures experienced by these fishes ranged from 1.7−25°C (endmembers are Coryphaenoides armatus and Platyccephalus indicus, respectively), with 17 species inhabiting waters >10°C (Fig. 1). See Fig. S2 in Supplement 3 for depth and temperature distributions for the 3 most populous orders in our data set. Average oxygen concentrations ranged from 0.5−7.0 ml l−1 (Anoplopora fimbria and Hippoglossus hippoglossus, respectively; Table 1). Estimated Lutz POC fluxes to yr−1 (Table 1). SCOC rates spanned from 0.45−1.7−25°C (endmembers are Coryphaenoides armatus and Platyccephalus indicus, respectively), with 17 species inhabiting waters >10°C (Fig. 1).

3.1. Pairwise GLM comparisons: growth rates have strong relationships with temperature, depth, and metabolic rate

As pairwise comparisons with environmental predictors were primarily used to inform which predictors should be included in the mixed-model analysis and may be highly susceptible to covariation, the results are not discussed at length here (see Figs. S3−S5 & S16 in Supplement 3). We do report results of pairwise comparisons with metabolic rates, as although these data were insufficient to include in the full mixed-model analysis, we found stronger correlations with depth than expected.

Metabolic rate positively correlated with metrics of growth. K and KD,Y increased with CS activity, both measured at 10°C and adjusted to each species’ habitat temperature (R² = 17−32%, p < 0.05; Fig. 4). CS activity adjusted to prematurity temperatures explained the greatest amount of variation among the metabolic rate predictors (R² = 27%, p < 0.05 with C. armatus, 32% without C. armatus). Beryx splendens was a notable outlier in metabolic rate relative to all metrics of growth (Fig. 4). When B. splendens was omitted from these analyses, metabolic rate was a stronger predictor of K (p < 0.05 for all metabolic rate metrics, R² = 20, 36, and 41% for CS10, lifespan CS10, and prematurity CS10, respectively) and KD,Y (p < 0.01 for all metabolic rate metrics, R² = 30, 40, and 48% for CS10, lifespan CS10, and prematurity CS10, respectively). CS activities were not significantly related to AIM or AIMD,Y unless B. splendens was removed from the analysis, in which case there were significant positive relationships (p < 0.001, R² = 37 and 34% for AIM vs. lifespan CS10 and prematurity CS10, respectively; p < 0.01, R² = 29, 42, and 37% for AIMD,Y vs. CS10, lifespan CS10, and prematurity CS10, respectively; Fig. 4).
Table 1. Depths, habitat variables, and growth metrics for each fish population (see Methods 2.1 and Supplement 2 for sources and explanation of depth ranges). $K$: von Bertalanffy growth coefficient calculated with length vs. calendar age in years. $K_{\text{GDY}}$: von Bertalanffy growth coefficient calculated with length vs. thermal age in growing degree-years. AIM: average increase in fish mass between calendar years during the prematurity growth phase. $A_{\text{IMDY}}$: average increase in fish mass between thermal growing degree-years during the prematurity growth phase. $A_M$: age at 50% maturity unless otherwise noted with the following subscripts: a: median of known minimum and maximum ages at maturity; b: estimated using a published length at 50% maturity and the von Bertalanffy curve; c: general time at maturity. Lutz particulate organic carbon (POC) flux is reported as 3 values separated left to right by slashes as values at minimum, median, and maximum depths; sediment community oxygen consumption (SCOC) values are below POC flux in *italics*. CS$_{HC}$: citrate synthase activity (units per gram wet weight of white muscle tissue at each species' average prematurity temperature). Ontogenetic migrators are underlined, and the prematurity version of applicable predictors are in square brackets. *Sebastes mentella* is not classified as an ontogenetic migrant here, but sampling was performed at 2 locations with slightly different temperatures, and so has a different prematurity temperature. Asterisks denote where there was insufficient data to calculate growth metrics; dashes indicate data not available. Number of fish represented in data from each study listed under n. Only observations containing both age and size are counted here. Data for *M. productus* contained many more observations of mass-at-age than of length-at-age, denoted with (L) for length observations and (M) for mass.

| Beryciformes | | | Depth | Lutz POC flux | O$_2$ | T | $K$ | $K_{\text{GDY}}$ | AIM | $A_{\text{IMDY}}$ | CS$_{HC}$ | $A_M$ | n |
|--------------|------------------|----------|-------|------------------|-------|-------|-------|-------|-------|------------------|-----------|-------|-----|
| *Beryx decadactlyus* | Charleston Bump | (Friess & Sedberry 2011) | 368 | 432 | 496 | 7.7/7.3/7.0 | 4 | 15.2 | 0.12 | 0.007 | * | * | – | 4 | 155 |
| *Beryx splendens* | Palliser Bank, NZ (NIWA) | | 200 | 500 | 800 | 33/26/21 | 5 | 9 | 0.22 | 0.021 | 33.1 | 2.34 | 3.41 | 5 b | 2310 |
| *Hoplostethus atlanticus* | Walter's Shoal, SW Indian Ocean (NIWA) | | 1000 | 1125 | 1250 | 6.9/6.4/5.9 | 4.5 | 5.4 | 0.06 | 0.007 | * | * | 0.5 | 22 | 409 |
| *Hoplostethus mediterraneus* | Eastern Mediterranean | (D’Onghia et al. 1998) | 388 | 642.5 | 897 | 8.2/7.0/6.0 | 4.4 | 13.9 | 0.21 | 0.013 | 72.7 | 4.33 | – | 4 | 419 |
| Gadiformes | | | | | | | | | | | | | | |
| *Antimora rostrata* | North Atlantic | (Orlov et al. 2018) | 1300 | 1900 | 2500 | 9.5 | 6.6 | 3.5 | 0.06 | 0.007 | 24.2 | 3.2 | 0.21 | 16 | 340 |
| *Coelorinchus caelorhincus* | Eastern Mediterranean | (Labropoulou & Papaconstantinou 2000) | 350 | 450 | 550 | 8.4/7.9/7.4 | 4.5 | 14.1 | 0.05 | 0.003 | 116.3 | 4.58 | – | 6 | 244 |
| *Coryphaenoides acrolepis* | West coast USA (Matsui et al. 1990, Andrews et al. 1999) | | 700 | 1350 | 2000 | 18/11/7.5 | 2.4 | 0.9 | 3.1 | 0.05 | 0.007 | 9.7 | 1.5 | 0.51 | 36 a | 146 |
| *Coryphaenoides armatus* | Mid-Atlantic Ridge | (Bergstad 2013) | 2000 | 3000 | 4000 | 6.9/4.0/2.8 | 6.3 | 2.9 | 0.04 | 0.007 | 36.5 | 5.23 | 1.04 | 20 c | 317 |
| | Station M | (Gerringer et al. 2018) | 2000 | 3150 | 4300 | 6.3/4.3/3.5 | 2.7 | 1.7 | * | * | 15.3 | 3.04 | 1.04 | 20 c | 107 |
| *Coryphaenoides rupestris* | British Isles | (Allain & Lorance 2000) | 400 | 950 | 1500 | 34.4/21/13 | 5.5 | 7.6 | 0.04 | 0.004 | 17.1 | 1.48 | 0.3 | 10 | 2814 |
| *Gadus morhua* | North Atlantic | (Rideout et al. 2016) | 50 | 125 | 200 | 34/32/30 | 6.2 | 3.2 | 0.13 | 0.022 | 40 | 6.47 | – | 5 | 7781062 |
| *Hymenocephalus italicus* | Eastern Mediterranean | (D’Onghia et al. 2000) | 305 | 601 | 897 | 9.7/8.0/6.6 | 4.4 | 13.9 | 0.18 | 0.011 | 78 | 3.46 | – | 3 | 408 |

Table continued on next page
## Table 1 (continued)

| Species                          | Location                        | Min. Depth (m) | Median Depth (m) | Max. Depth (m) | Lutz POC flux (mg C m⁻² d⁻¹) | SCOC (O₂ m⁻² d⁻¹) | O₂ (ml l⁻¹) | T (°C) | K     | Kᵦ⁻¹ | AIM (%) | AIMᵦ⁻¹ (%) | CSₑ, n  |
|----------------------------------|---------------------------------|----------------|------------------|----------------|------------------------------|-------------------|-------------|--------|-------|-------|--------|-------------|--------|
| *Lepidion eques*                 | North Atlantic                  | 500            | 700              | 900            | 31/26/21                     | 6.1               | 5.8         | 0.1    | 0.006 | 43.9  | 3.99   | –          | 9      | 295      |
|                                  | (Magnússon 2001)                |                |                  |                |                              |                   |             |        |       |       |        |             |        |          |
| *Macrourus carinatus*            | Heard Island                    | 500            | 750              | 1000           | 11/8.7/6.9                   | 4.4               | 2.3         | 0.07   | 0.013 | 19.3  | 3.43   | –          | 11     | 142      |
|                                  | (Van Wijk et al. 2003)          |                |                  |                |                              |                   |             |        |       |       |        |             |        |          |
| *Macrouronus novaebraelandiae*   | Off New Zealand                 | 200            | 500              | 800            | 39/31/25                     | 5.1               | 9.2         | 0.17   | 0.014 | 43    | 2.94   | –          | 6      | 9569     |
|                                  | (Horn & Sullivan 1996)          |                |                  |                |                              |                   |             |        |       |       |        |             |        |          |
| *Merluccius australis*           | West Coast South Island         | 300            | 700              | 1100           | 25/18/14                     | 4.6               | 7.7         | 0.22   | 0.02  | 49.6  | 3.6    | –          | 6      | 3884     |
|                                  | (NIWA)                          |                |                  |                |                              |                   |             |        |       |       |        |             |        |          |
| *Merluccius hubbsi*              | Off Brazil                      | 100            | 150              | 200            | 16/16/15                     | 5                 | 18          | 0.19   | 0.009 | *     | 5.67   | –          | 2      | 685      |
|                                  | (Vaz-dos-Santos et al. 2007)    |                |                  |                |                              |                   |             |        |       |       |        |             |        |          |
| *Merluccius merluccius*          | Iberian Peninsula               | 70             | 220              | 370            | 41/36/31                     | 6.6               | 12.6        | 0.1    | 0.007 | *     | *     | –          | 3      | 965      |
|                                  | (Pitieiro & Sainza 2003)        |                |                  |                |                              |                   |             |        |       |       |        |             |        |          |
| *Merluccius productus*           | Northern West Coast USA         | 50             | 275              | 500            | 67 [67]                       | 3.1               | 7.1         | 0.27   | 0.025 | 122.6 | 7.22   | 1.52       | 3      | 5501 [M] |
|                                  | (NWFSC Groundfish Trawl Survey) | [120]          | [170]            | [220]          | 7.3 [7.7]                     | [3.0]             | [7.2]       |        |       |       |        |             |        |          |
|                                  | Southern West Coast USA         | 50             | 275              | 500            | 36 [37]                       | 2.4               | 7.5         | 0.2    | 0.019 | 88.3  | 5.91   | 1.65       | 4475   | [M]      |
|                                  | (NWFSC Groundfish Trawl Survey) | [170]          | [190]            | [210]          | 11 [4.4]                      | [2.6]             | [7.9]       |        |       |       |        |             |        |          |
| *Micromesistius australis*       | Off Argentina                   | 70             | 200              | 620            | 64/56/37                      | 6.4               | 4.5         | 0.17   | 0.023 | 81.2  | 8.16   | –          | 4      | 4919     |
|                                  | (Cassia 2000)                   |                |                  |                |                              |                   |             |        |       |       |        |             |        |          |
|                                  | Off New Zealand                 | 400            | 500              | 600            | 13/12/11                      | 5.8               | 5.8         | 0.2    | 0.022 | 96.7  | 7.76   | –          | 3      | 7920     |
|                                  | (NIWA)                          |                |                  |                |                              |                   |             |        |       |       |        |             |        |          |
| *Mora moro*                     | Chatham Rise                    | 500            | 750              | 1000           | 23/18/15                      | 5                 | 6.3         | 0.13   | 0.014 | 42.8  | 3.81   | 0.34       | 8      | 294      |
|                                  | (Sutton et al. 2010)            |                |                  |                |                              |                   |             |        |       |       |        |             |        |          |
| *Nezumia sclerorhynchus*         | Eastern Mediterranean           | 350            | 550              | 750            | 8.4/7.4/6.6                   | 4.5               | 14.1        | 0.15   | 0.009 | 42.5  | 1.97   | –          | 11     | 200      |
|                                  | (Labropoulou & Papaconstantinou 2000) |            |                  |                |                              |                   |             |        |       |       |        |             |        |          |
| *Trachyrhynchus scabrus*         | Western Mediterranean           | 700            | 875              | 1050           | 11/9.4/8.4                    | 4.3               | 13.3        | 0.15   | 0.009 | *     | *     | –          | 4      | 263      |
|                                  | (Massutí et al. 1995)           |                |                  |                |                              |                   |             |        |       |       |        |             |        |          |

**Ophidiiformes**

*Genypterus blacodes*  
Campbell Plateau, NZ (NIWA)  
200 500 800 13/10/8 5.8 6.6 0.15 0.015 65.7 5.31 – 6 5254  
Chatham Rise, NZ (NIWA)  
200 500 800 28/22/17 4.2 5.5 7.3 0.12 0.011 66.8 5.05 – 6 5751

**Perciformes**

*Cheilodactylus spectabilis*  
Tasmania (Ewing et al. 2007)  
0 25 50 34/33/33 5.6 15.3 0.39 0.026 57.7 2.54 – 3 198
### Table 1 (continued)

| Species                  | Location Description | Min. Depth (m) | Median Depth (m) | Max. Depth (m) | Lutz POC flux (mg C m\(^{-2}\) d\(^{-1}\)) | SCOC (O\(_2\) m\(^{-2}\) d\(^{-1}\)) | O\(_2\) (ml l\(^{-1}\)) | T (°C) | K | K\(_{CDY}\) | AIM (%) | AIM\(_{CDY}\) (%) | CS\(_I\) | A\(_M\) | n |
|--------------------------|----------------------|----------------|------------------|---------------|---------------------------------------------|-------------------------------------|--------------------------|--------|---|----------------|--------|-------------------|--------|--------|---|
| *Epigonus telescopus*    | Off New Zealand      | 500            | 800              | 1000          | 15/12/10                                    | 4.8                                | 7.2                      | 0.04   | 0.004 | 5.7           | 0.54   | –                 | 36     | 554    |   |
| *Hyperoglyphe perciformis* | Charleston Bump     | 200            | 363              | 526           | 10/8.8/7.8                                   | 4.1                                | 16.1                     | 0.21   | 0.012 | 9.2           | 0.46   | –                 | 6      | 710    |   |
| **Pleuronectiformes**    |                      |                |                  |               |                                             |                                    |                          |        |    |                |        |                   |        |        |   |
| *Atheresthes stomias*    | West coast USA       | 27             | 148.5            | 270           | 51 [55]                                      | 1.4                                | 6.4                      | 0.14   | 0.012 | 33.2          | 2.68   | 0.46              | 7      | 3525   |   |
|                          | (NWFSC Groundfish    | [130]          | [170]            | [210]         | 7.7 [7.6]                                    | [1.7]                              | [7.6]                    |        |      |                |        |                   |        |        |   |
|                          | Trawl Survey)        |                |                  |               |                                             |                                    |                          |        |    |                |        |                   |        |        |   |
| *Cynoglossus zanzibarensis* | South Africa       | 30             | 230              | 430           | 59/50/43                                     | 4.7                                | 14.8                     | 0.44   | 0.025 | 100.6         | 4.05   | –                 | 2      | 180    |   |
|                          |                      |                |                  |               |                                             |                                    |                          |        |    |                |        |                   |        |        |   |
| *Eopsetta jordani*       | West coast USA       | 160            | 205              | 250           | 46 [51]                                      | 3.1                                | 8.4                      | 0.13   | 0.009 | 39.7          | 2.75   | 0.64              | 5      | 4745   |   |
|                          | (NWFSC Groundfish    | [80]           | [100]            | [120]         | 7.0 [9.2]                                    | [4.1]                              | [10.6]                   |        |      |                |        |                   |        |        |   |
|                          | Trawl Survey)        |                |                  |               |                                             |                                    |                          |        |    |                |        |                   |        |        |   |
| *Hippoglossus hippoglossus* | North Atlantic    | 55             | 81.5             | 108           | 47/45/44                                     | 7                                  | 6.2                      | 0.13   | 0.014 | 186.1         | 6.69   | –                 | 6      | 2429   |   |
|                          |                      |                |                  |               |                                             |                                    |                          |        |    |                |        |                   |        |        |   |
| *Microstomus pacificus*  | West coast USA       | 55             | 777.5            | 1500          | 31 [43]                                      | 1.1                                | 5.3                      | 0.17   | 0.017 | 23.3          | 2.1    | 0.78              | 7      | 6020   |   |
|                          | (NWFSC Groundfish    | [110]          | [180]            | [250]         | 5.4 [8.2]                                    | [3.3]                              | [8.8]                    |        |      |                |        |                   |        |        |   |
|                          | Trawl Survey)        |                |                  |               |                                             |                                    |                          |        |    |                |        |                   |        |        |   |
| *Parophrys vetulus*      | West coast USA       | 40             | 120              | 200           | 49 [51]                                      | 3.8                                | 9.2                      | 0.35   | 0.028 | 61.6          | 3.31   | 0.85              | 4\(_s\) | 772    |   |
|                          | (NWFSC Groundfish    | [70]           | [90]             | [110]         | 11 [6.4]                                     | [4.2]                              | [10.6]                   |        |      |                |        |                   |        |        |   |
|                          | Trawl Survey)        |                |                  |               |                                             |                                    |                          |        |    |                |        |                   |        |        |   |
| *Reinhardtius hippoglossoides* | North Atlantic | 500            | 750              | 1000          | 23/18/14                                     | 6.7                                | 2                        | 0.07   | 0.015 | 58.4          | 9.78   | –                 | 10     | 266    |   |
|                          | (Dwyer et al. 2016)  |                |                  |               |                                             |                                    |                          |        |    |                |        |                   |        |        |   |
| **Scorpaeniformes**      |                      |                |                  |               |                                             |                                    |                          |        |    |                |        |                   |        |        |   |
| *Anoplopoma fimbria*     | West coast USA       | 200            | 600              | 1000          | 26 [40]                                      | 0.7                                | 4.3                      | 0.43   | 0.045 | 61.4          | 4.87   | 1.67              | 5      | 3772   |   |
|                          | (NWFSC Groundfish    | [140]          | [270]            | [390]         | 4.4 [6.2]                                    | [2.7]                              | [7.4]                    |        |      |                |        |                   |        |        |   |
|                          | Trawl Survey)        |                |                  |               |                                             |                                    |                          |        |    |                |        |                   |        |        |   |
| *Chelidonichthys capensis* | South Africa       | 30             | 115              | 200           | 42/39/37                                     | 4.7                                | 14.8                     | 0.07   | 0.004 | 75.2          | 3.31   | –                 | 4      | 383    |   |
|                          |                      |                |                  |               |                                             |                                    |                          |        |    |                |        |                   |        |        |   |
| *Chelidonichthys queketti* | South Africa       | 50             | 100              | 150           | 38/36/35                                     | 4.7                                | 14.8                     | 0.4    | 0.022 | *             | *      | –                 | 2      | 227    |   |
|                          |                      |                |                  |               |                                             |                                    |                          |        |    |                |        |                   |        |        |   |
| *Helicolenus percoides*  | Off New Zealand     | 100            | 400              | 700           | 37/29/22                                     | 5.4                                | 8.8                      | 0.1    | 0.009 | 53.9          | 3.73   | –                 | 5      | 1338   |   |
|                          | (Paul & Horn 2009)   |                |                  |               |                                             |                                    |                          |        |    |                |        |                   |        |        |   |
| *Platycephalus indicus*  | Japan                | 0              | 15               | 30            | 12/12/2012                                   | 4.8                                | 25                       | 0.38   | 0.014 | 118.7         | 2.78   | –                 | 2      | 351    |   |

*Table continued on next page*
| Species                  | Location                          | Min. Depth (m) | Median Depth (m) | Max. Depth (m) | Lutz POC flux (mg C m$^{-2}$ d$^{-1}$) | SCOC (O$_2$ m$^{-2}$ d$^{-1}$) | O$_2$ (ml l$^{-1}$) | T (°C) | K | K$_{GDY}$ | AIM (%) | AIM$_{GDY}$ (%) | CS$_H$ | A$_M$ | n   |
|--------------------------|-----------------------------------|----------------|------------------|----------------|----------------------------------------|-------------------------------|-------------------|--------|---|----------|--------|------------------|-------|-------|-----|
| *Pterois volitans*       | Gulf of Mexico                    | 50             | 85               | 120            | 24/23/22/13                             | 4.2                           | 20.8              | 0.44   | 0.018 | 163.5       | 5.92   | –                 | 1     | 1576  |
| *Scorpaena notata*       | Eastern Mediterranean             | 30             | 365              | 700            | 51/38/29/13                             | 5.1                           | 13.4              | 0.32   | 0.019 | 59           | 2.85   | –                 | 2.5   | 225   |
| *Sebastes crameri*       | West coast USA (Nichol 1990)      | 140            | 175              | 210            | 73/71/69/13                             | 2.9                           | 7.6               | 0.2    | 0.019 | 75.9         | 5.66   | 0.49              | 5     | 1060  |
| *Sebastes ensifer*       | West coast USA (Love et al. 2018) | 45             | 205.5            | 366            | 62/54/47/13                             | 3.1                           | 8.7               | 0.13   | 0.012 | 23.7         | 1.91   | –                 | 9     | 171   |
| *Sebastes fasciatus*     | Flemish Cap (Saborido-Rey et al. 2004) | 130            | 315              | 500            | 34/28/24/13                             | 6.3                           | 3.9               | 0.13   | 0.018 | *           | *      | –                 | 6     | 5174  |
| *Sebastes marinus*       | Flemish Cap (Saborido-Rey et al. 2004) | 100            | 200              | 300            | 35/32/29/13                             | 6.3                           | 3.9               | 0.07   | 0.009 | 52.1         | 6.15   | 0.32              | 34    | 332   |
| *Sebastes melanostomus*  | West coast USA (Stevens et al. 2004) | 250            | 425              | 600            | 40/35/30/13                             | 1                             | 6.4               | 0.06   | 0.007 | 7.9          | 0.82   | 0.32              | 34    | 332   |
| *Sebastes mentella*      | Irminger (Stranksy et al. 2005)   | 300            | 600              | 900            | 29/22/16/13                             | 6.7                           | 3.8               | 0.13   | 0.02  | 10.2         | 1.58   | –                 | 10    | 359   |
| *Sebastes mystinus*      | Newfoundland (Saborido-Rey et al. 2004) | 300            | 600              | 900            | 29/22/17/13                             | 6.5                           | 3.7               | 0.11   | 0.016 | 27.8         | 3.64   | –                 | 10    | 8930  |
| *Sebastolobus alascanus* | Off California (Kline 1996)        | 90             | 750              | 1500           | 31/42/13/13                             | 0.5                           | 4.4               | 0.02   | 0.002 | 21.8         | 2.26   | 0.4               | 12    | 1140  |
| *Sebastolobus altivelis* | Off California (Kline 1996)        | 400            | 950              | 1500           | 47/30/21/13                             | 0.7                           | 4.2               | 0.06   | 0.008 | *           | *      | 0.21              | 6c    | 342   |
| *Zeiformes*              | *Allocyttus niger*                | 700            | 900              | 1300           | 17/18/13/13                             | 4.8                           | 3.9               | 0.23   | 0.023 | 3.7          | 0.54   | –                 | 27c   | 2056  |
|                          | (NIWA)                            | [644]          | [840]            | [1039]         | 2.7                                      | [4.9]                         | [4.2]             |        |    |          |        |                  |       |       |
|                          | *Allocyttus verrucosus*           | 1000           | 1200             | 1400           | 11/10/8.8/13                            | 4.2                           | 3.6               | 0.06   | 0.01  | 12.4         | 1.79   | –                 | 24    | 102   |
|                          | Off Tasmania (Stewart et al. 1995) | 600            | 700              | 800            | 17/15/14/13                             | 5.2                           | 7.6               | 0.05   | 0.005 | 8.5          | 0.77   | –                 | 33b   | 97    |

Table 1 (continued)
3.2. Multivariate GLMM analysis

The GLMM for prematurity analysis explained more variation in all growth metrics than that for lifespan. Therefore, we report only the prematurity GLMM analysis here. See Figs. S9 & S11, Tables S7 & S8 in Supplement 3 for GLMM results upon excluding *C. armatus*. *K* was best explained by a negative relationship with minimum depth and a positive relationship with temperature ($R^2 = 47\%$, AICc = −127.9, hierarchical partitioning: 71\% for depth, 28\% for temperature, predictors removed: NPP; Table 2, Fig. 5). Five additional models were just as likely as the best model by ΔAICc < 2, including some combination of minimum depth, temperature, Lutz POC flux, and locomotory mode (Table 2). The second-best model indicated that *K* was nearly just as likely to be explained by depth alone ($R^2 = 45\%$, AICc = −127.7, predictors removed: NPP). Random-effect intercepts for each phylogenetic family all showed overlapping confidence intervals, with sebastids appearing slightly farther from the mean (Fig. S6 in Supplement 3). This is likely due to the large variability of growth rate within the family and the especially slow growth of *Sebastolobus alascanus*.

**Fig. 1.** Numbers of populations of fishes in this study organized by (A) lifespan depth category and (B) average temperature of habitat. Median depths across the entire fish lifespan were used to categorize each species.

**Fig. 2.** World map highlighted with catch locations from each study used for the meta-analysis. Catch locations are colored by Order.
When a random effect for phylogenetic family was not included, locomotory mode was also included, along with depth and temperature, as an important predictor in the best model, with benthopelagic fishes growing more quickly than benthic fishes (Table S2 in Supplement 3). However, hierarchical partitioning revealed that depth and temperature were by far the primary drivers ($R^2 = 49\%$, weight = 0.36, hierarchical partitioning: 64\% for depth, 30\% for temperature, and 6\% for locomotory mode, predictors removed: NPP). Rerunning the best model with the addition of prematurity CS$_{11}$ as a predictor increased $R^2$ to 61\% but limited the sample size to 19 and increased the maximum VIF of the predictors to 2.64.

$K_{GDY}$ was best explained by minimum depth and locomotory mode, showing declines with depth and in benthic fishes in the GLMM analysis ($R^2 = 27\%$, AICc = $-399.2$, hierarchical partitioning: 83\% for depth, 17\% for locomotory mode, weight = 0.64, predictors removed: NPP; Table 3, Fig. 6). Only one additional model was equally likely (i.e. $\Delta$AICc < 2), and contained POC flux, minimum depth, and locomotory mode as important predictors ($R^2 = 29\%$, AICc = $-398$, predictors removed: NPP; Table 3). No changes in important predictors chosen occurred when phylogenetic family was not included in the analysis as a random effect. Random-effect intercepts for each phylogenetic family all showed overlapping confidence intervals (Fig. S7 in Supplement 3). Rerunning...
Black et al.: Drivers of growth in deep-sea teleosts

this best model with prematurity CSH as a predictor increased $R^2$ to 47% but limited the sample size to 19.

AIM was best explained by temperature and depth, but the importance of depth was sensitive to C. armatus inclusion. When C. armatus was included, GLMM analysis indicated that variation in AIM is best explained by temperature, Lutz POC flux, and $O_2$, in that rank order ($R^2 = 38\%$, weight = 0.37, hierarchical partitioning: 72% for depth, 20% for temperature, and 8% for $O_2$; predictors removed: NPP). There were again 3 additional models with $\Delta$AICc < 2, each containing some combination of these same predictors ($R^2 = 33-40\%$, predictors removed: NPP; Table S3 in Supplement 3). Re-running this best model with prematurity CSH as a predictor increased $R^2$ slightly to 40% but limited sample size to 18, prevented the use of phylogeny as a random effect, and the maximum VIP of the predictors increased to 3.11.

AIMGDY declined with decreasing $O_2$, increasing depth, and in benthic fishes, but results are sensitive to C. armatus inclusion. When C. armatus was included in the analysis, AIMGDY was found to be best explained by $O_2$ and Lutz POC flux ($R^2 = 22\%$, AICc = 73.3, weight = 0.40, hierarchical partitioning: 80% for $O_2$, 20% for Lutz POC flux, predictors removed: NPP; Table 5, Fig. 8). Two additional models were equally likely by $\Delta$AICc < 2: one model containing $O_2$ alone ($R^2 = 18\%$, AICc = 73.3, weight = 0.40; Table 5) and another model containing $O_2$, locomotory mode, and Lutz POC flux ($R^2 = 24\%$, AICc = 74.7, weight = 0.20; Table 5). Random-effect intercepts for each phylogenetic family all showed overlapping confidence intervals, but there are a few apparent outliers in the model residuals (Fig. S10 in Supplement 3). When random effects were excluded, these patterns were similar, but with a drop in $R^2$ for one of the 3 output models ($R^2 = 13-20\%$; Table S6 in Supplement 3). Re-running this best model with prematurity CSH as a predictor increased $R^2$ to 32% but limited sample size to 18.

4. DISCUSSION

This study presents the first comprehensive examination of the environmental and physiological drivers of demersal fish growth rates across the depth continuum. The general decline in the von Bertalanffy growth coefficient $K$ with depth observed by Drazen & Haedrich (2012) was found again here (Fig. 5). Prematurity depths for ontoge-
netic migrators explained more variation in $K$ and $K_{GDY}$ than when lifespan-averaged depths were used (Table S1 in Supplement 3). This is not surprising, given that most fish growth occurs in the prematurity phase of life and thus habitat variables must be matched accordingly. Average percent increase in mass during the prematurity phase (i.e. AIM) also declined with depth, though this was sensitive to the inclusion of *Coryphaenoides armatus* (Table S1 in Supplement 3, Fig. 7). However, observing the depth pattern alone does not lead to an understanding of the underlying drivers because depth is not a direct ecological driver, and there is still considerable variation in growth (~60–90%, depending on the metric) that is unexplained by depth alone (Table S1 in Supplement 3).

Temperature during the prematurity phase was found to be an important predictor of $K$ and AIM (Table S1 in Supplement 3, Figs. 5 & 7). Our findings reinforce the exponential increases in interspecific growth rate as a function of temperature hypothesized in prior work (Van der Have & De Jong 1996, Jobling 1997, Neuheimer & Grønkjær 2012, Rall et al. 2012). It should be noted that the patterns with temperature in this interspecific meta-analysis do not reflect those found in intra-specific studies. While populations of a given species in warmer temperatures have been shown to grow faster than conspecifics inhabiting colder temperatures, this intraspecific growth response is unimodal: growth rate initially increases with temperature to an optimum and decreases beyond this threshold (Braaten & Guy 2002, Pörtner & Knust 2007, Thresher et al. 2007, Neuheimer et al. 2011, Rypel 2012). A similar meta-analysis of growth in freshwater fishes across South America by de Santana et al. (2020) primarily focused on the effects of latitude as a proxy for temperature but did not find a significant relationship with interspecific growth rate, although they found that intraspecific asymptotic length decreases with latitude. Their meta-analysis differed in that it focused on freshwater habitat, used latitude as a proxy for temperature (which is imprecise), and incorporated ageing methods not included here (scales, vertebrae, length–frequency analysis, etc.).

Contrary to expectations, the apparent influence of depth was still present even after the signal of temperature was accounted for with temperature-adjusted growth rates. Multivariate mixed models showed that 26% of the remaining variability in $K_{GDY}$ is explained by depth and locomotory mode. This would suggest that temperature is not the sole driver of variation in $K$ across depth. Compared to patterns seen with $K_{GDY}$, depth was found to be a lesser, but still important, predictor of variation in AIM$_{GDY}$, though only upon exclusion of *C. armatus* from the data set. The sensitivity of these patterns in AIM and AIM$_{GDY}$ to inclusion of *C. armatus* likely reflects inherent uncertainty as to this species’ age-at-matu-

| Models | Int. Loc. Min. Lutz POC flux Temp. | $R^2$ | ΔAICc | Weight |
|--------|--------|--------|--------|--------|
| 1      | 7.02   | 3.345  | -0.6743 | 0.472  | 0.205  |
| 2      | 6.846  | 3.881  | -0.7151 | 0.4916 | 0.267  |
| 3      | 8.049  | +      | 3.61    | -0.7975| 0.5144 | 0.451  |
| 4      | 8.149  | +      | 2.835   | -0.7975| -1.027 | 0.5144 | 0.164  |
| 5      | 6.977  | +      | 2.641   | -0.7279| -0.9488| 0.49   | 0.154  |
| 6      | 7.825  | +      | 4.144   | -0.7279| -0.9488| 0.4631 | 0.129  |

Fig. 5. Effects of the top mixed model of $K$ vs. minimum depth when temperature is held at the median (green, dashed line), first quartile (yellow, dotted line), and third quartile (blue, dashed line). Rug values for each data point are colored according to closest proximity to median, first quartile, or third quartile.
rity, as well as a lack of data for other abyssal species, which is discussed in more detail below. Nevertheless, the persistent significance of depth in explaining growth suggests the influence of additional variables associated with depth in determining growth.

The estimates of food availability used here were not found to be important predictors of interspecific growth rates (Table 2). Lutz POC flux was a secondary predictor of AIM and AIMDGY, but only when the outlier *C. armatus* was included. These results echo the findings of the meta-analysis by McClain et al. (2012) in which the amount of chemical energy present in the base of the food web was not found to control interspecific patterns in rockfish growth rates, but instead constrains population- or ecosystem-level processes such as abundance, biodiversity, and total biomass. Indeed, fisheries research has found that wild populations increase in abundance until reaching a carrying capacity determined at least in part by availability of prey (reviewed in Anderson 1988 and Le Pape & Bonhommeau 2015). Population sizes of abyssal fishes are thought to track long-term fluctuations in prey abundance, with interannual changes in population densities likely due to migration rather than increased reproduction rates (Drazen et al. 2012, Milligan et al. 2020). Studies on the energetics of individual growth in fishes have established that an increase in food ration leads to more rapid growth, and intraspecific variation in growth rates on short timescales, such as in aquaculture settings, clearly affects fish growth rates (Persson & De Roos 2006). Despite food's importance as an intraspecific driver of individual growth in aquaculture settings, it appears that interspecific patterns in growth across depth are relatively unaffected by variability in food supply as measured via Lutz POC, NPP, or SCOC. However, we were unable to incorporate these 2 predictors into GLMM analysis due to covariation and lack of data, respectively. It is possible, given more data on SCOC rates or other food supply metrics, that a significant relationship with growth would be apparent.

Oxygen was found to be an important predictor of AIM, though relatively less so than depth and temperature. In GLMM analyses, O₂ only explained about 5% of variation in AIM. After correcting for temperature, the influence of O₂ was much stronger, becoming the most important predictor of AIMDGY with or without the inclusion of *C. armatus*. However, this pattern with AIMDGY was driven by several oxygen minimum zone (OMZ)-dwelling fishes inhabiting O₂ concentrations of less than 2 ml l⁻¹. Oxygen has been linked with growth in the literature: overall population biomass, maximum body size, and abundance tends to correlate with oxygen levels within OMZs (reviewed in Gallo & Levin 2016). Intraspecific studies have shown a link between metabolism, oxygen concentration, and growth, with hypoxic conditions suppressing growth and metabolic rate in actively swimming fishes, though the critical concentration at which suppression occurs is species-specific (reviewed in Gray et al. 2002). Moreover, the heightened longevity observed in deeper dwelling fishes has been attributed to lower concentrations of
oxygen and thus free radical damage to cells, though high longevity is not necessarily coupled to growth rate (Cailliet et al. 2001). Generally, less is known regarding the effects of oxygen on interspecific growth rates. High O2 demand has been associated with rapidly growing, ‘high performance’ fishes such as tuna and mahi mahi (Brill 1996). However, prior work on demersal species has not shown a clear pattern between environmental O2 concentration and rate of metabolism. Rather, a strong positive correlation between high metabolism and mass-specific gill surface area has been found (Friedman et al. 2012), and deep-sea species appear to have scaling exponents of metabolic rate that indicate they are less limited by geometric constraints of oxygen transport than shallower fish (Killen et al. 2010). Our analysis shows that sablefish live at low concentrations of oxygen (~0.7–2.7 ml l−1) and have high rates of metabolism and growth, but also were found by Friedman et al. (2012) to have relatively higher gill surface area than several of the slow-growing OMZ-dwellers in our data set, such as Sebastes alascanus, S. altivelis, and Microstomus pacificus. The positive relationship between growth and O2 we found was driven entirely by these slow-growing OMZ-dwellers, which may indicate a threshold response in that oxygen only limits maximum growth rate at concentrations lower than about 2 ml l−1, and only in fishes without specific adaptations for high gill surface area.

The influence of phylogeny on growth appears to be minor, though we were unable to conclusively and quantitatively determine its effect on growth. Across phylogenetic orders, no significant differences were found in $K$ or $K_{GDY}$. Only Zeiformes (the oreos) were found to grow significantly slower than other orders in terms of AIM and AIM$_{GDY}$. This may be due simply to a lack of sampling in this order because all 3 species analyzed are deep living, with median lifespan

Table 4. Results of dredge analysis of AIM vs. prematurity predictors including Coryphaenoides armatus and with phylogenetic family as a random effect, showing models with a ΔAICc score of <2 relative to the top model. For numerical predictors, the intercept (Int.) is shown if the predictor is in the model. Global model call: glmer(formula = AIM ~ scale(med. depth) + scale(Lutz POC flux to max depth) + scale(O2) + scale(temperature) + locomotory mode + (1 | phyl. family), family = Gamma, na.action = 'na.fail'). Predictors removed: NPP

| Models | Int. | Loc. | Min. | Lutz | Temp | R² | AICc | ΔAICc | Weight |
|--------|------|------|------|------|------|----|------|-------|--------|
| 1      | 1.361| −0.3944| −0.3187| −0.3689| 0.3799| 81.2| 0 | 0.367 |
| 2      | 1.337| 0.5791| −0.2696| −0.1653| 0.3687| 82.1| 0.87| 0.237 |
| 3      | 1.377| 0.3014| −0.3561| −0.2198| −0.2738| 0.4012| 82.2| 1.02| 0.22 |
| 4      | 1.292| 0.7638| −0.2156| 0.3261| 0.3261| 82.7| 1.47| 0.176 |

Fig. 7. (A) Effects of the top mixed model of AIM vs. temperature when oxygen is held constant at the median (green, solid line), first quartile (blue, dashed line), and third quartile (yellow, dotted line) values. Particulate organic carbon (POC) flux is held constant at the median value. (B) Effects of the top mixed model of AIM vs. temperature when POC flux is held constant at the median (green, solid line), first quartile (blue, dashed line), and third quartile (yellow, dotted line) values, and oxygen concentration is held constant at the median value (4.8 ml l−1). Rug values for each data point are colored according to closest proximity to median, first quartile, or third quartile

Fig. 7. (A) Effects of the top mixed model of AIM vs. temperature when oxygen is held constant at the median (green, solid line), first quartile (blue, dashed line), and third quartile (yellow, dotted line) values. Particulate organic carbon (POC) flux is held constant at the median value. (B) Effects of the top mixed model of AIM vs. temperature when POC flux is held constant at the median (green, solid line), first quartile (blue, dashed line), and third quartile (yellow, dotted line) values, and oxygen concentration is held constant at the median value (4.8 ml l−1). Rug values for each data point are colored according to closest proximity to median, first quartile, or third quartile
depths >600 m. We had insufficient data to correct for variability in responses to our environmental variables at the species level using random effects but were able do so at the family level. Important predictors of variation in AIM change when phylogeny is incorporated as a random effect (see results for GLMM analysis and Tables S4 & S5 in Supplement 3), but these are small shifts from one covarying factor to another (e.g. Lutz POC flux vs. depth). The effect of phylogeny is difficult to address, particularly because growth data are not available for shallow-water members of some deep-sea families, such as the Oreosomatidae and Macrouridae (Drazen & Haedrich 2012, Priede & Froese 2013). We also included the shallow dwelling ophiiid Genypterus blacodes but were unable to find age and growth data for deep-sea members of this family. Similar meta-analyses on chondrichthyan growth rates have investigated the importance of phylogeny as a predictor: García et al. (2008) found correlations with growth rates and genus, and Rigby & Simpfendorfer (2015) found that after accounting for phylogeny, the influence of environmental conditions on growth was still significant. Overall, the effect of phylogeny is inconclusive, but its incorporation into these analyses as a random effect appears to account for some underlying variation in growth. Our findings reinforce those of Rigby & Simpfendorfer (2015), that while phylogeny can be important, other environmental drivers seem to be more so.

Temperature and depth fall out from our analyses as the primary, indirect drivers of growth, and in addition we found a strong positive relationship to metabolic rate, though it is not clear whether it is a driver or covarying energetic parameter across the depth continuum. All 3 CS activity predictors—non-adjusted, adjusted to average temperature, and adjusted to average prematurity temperature—were found to be significantly related to nearly all the growth metrics in this study (Fig. 4). While there was insufficient data to include metabolic rate as a predictor in the full mixed-model analysis, adding it to the top mixed model for each growth metric drastically increased the variation explained. Metabolic rate in demersal fishes has been found to decline with depth even after accounting for temperature and body size effects (Sullivan & Somero 1980, Drazen & Seibel 2007, Drazen et al. 2015). It is yet unclear whether the depth-related decline in metabolic rate is mechanistically driving the patterns observed in growth rate or simply correlates to decreased maintenance costs and activity levels associated with increasing depth (Killen et al. 2010). Careau et al. (2008) suggest 2 possible models for energy budgeting of growth and metabolism. In the allocation model, metabolic rate reflects the baseline energetic cost of cellular maintenance, and leftover energy is funneled into activity and/or growth (i.e. growth rate can be negatively correlated with metabolism). In the performance model, the metabolic rate represents the mechanistic rates of biochemical processes.
including channeling of energy into growth, resulting in positive correlations between growth and metabolic rates. Both models have found some support in intraspecific studies, with relationships between growth and metabolic rate changing across low, intermediate, and ad libitum food levels (Careau et al. 2008, Jobling 2010, Killen 2014, Auer et al. 2015). Our findings of positive correlations between growth and metabolic rates generally support the performance model. Our findings also align with a broad pattern of ‘fast vs. slow’ life history adaptations observed across and within animal taxa. That is, variation in life history parameters can be partly explained by a fast vs. slow dichotomy, with organisms with a relatively faster pace of life having more proactive behavior, short times-to-maturity, higher activity levels, slower lifespan, and more rapid growth than ‘slower’ organisms. (Mittelbach et al. 2014, Auer et al. 2018, Damsgård et al. 2019; reviewed in Réale et al. 2010).

Ultimately, the question is what evolutionary pressure is driving slow growth and slow metabolic rates in deep-sea fishes? Studies on catch-up growth in teleosts have demonstrated that these organisms typically grow at sub-maximal rates, but grow faster upon feeding after a period of starvation, to the detriment of their tissue quality and overall lifespan (Metcalfe et al. 2002, Alvarez & Metcalfe 2005, Hector & Nakagawa 2012). Taken together, these patterns would suggest that growth rate is not strictly bound to metabolic rate as a function of temperature and body mass, but rather the deep sea broadly selects for slower paces of life, typically low growth and metabolic rates and long lifespans and maturation times. Within the physiological constraints on growth rate imposed by environmental conditions, the remaining variation in growth rate is likely driven by the evolutionary fitness conferred by slower rates in deeper environments. The visual interactions hypothesis has suggested that high metabolic rates (and thus high locomotory capacity and response rate) are relatively advantageous in shallower, more illuminated habitats to facilitate prey pursuit and predator avoidance over larger distances (Childress 1995). As light decreases with depth to ~1000 m, the accompanying decrease in escape distances, visual detection range, and overall predator and prey densities favors a slower metabolic rate and perhaps a suite of life history strategies including slow growth and reactive rather than proactive behavior. Predation risk from marine mammals also drops sharply with depth (though some marine mammals forage deeper than 1000 m), which might result in lower mortality risk and relaxing selective pressure for both fast growth/early maturity and high-performance swimming ability in prey fishes. The importance of locomotory mode as an important predictor of growth in some of our models may also reflect the fast vs. slow dichotomy of life history strategies (Réale et al. 2010), as we found the lower activity benthic species to grow slightly slower than more active benthopelagics. Interestingly, Beryx splendens, a deep seamount dweller with rapid metabolism, high activity, but a very slow growth rate, was an exception to our overall findings. This result matches earlier work suggesting that high metabolic rate but slow growth might be common in seamount fishes (Koslow 1996). In this case, it is possible that the seamount ecosystem, which is more swept by currents and attracts large numbers of predators (Morato et al. 2010) in relation to many other deep habitats, favors higher swimming and metabolic activity (Koslow 1996), but why this then favors slower growth (given the potentially higher predation risk) is unclear. Further, Hoplostethus atlanticus, another seamount species in our data set, had low growth but did not have a high metabolic rate. We suggest that further expanding both metabolic and growth rate measurements on seamount-associated fishes may be a way to more clearly identify evolutionary pressures on energetic parameters in deep-sea fishes.

The strong relationships found in this meta-analysis provide an essential first step in understanding the drivers of growth variation and in locating gaps in the current knowledge for targeted research. However, a fully predictive model of growth across the depth continuum or a complete mechanistic understanding of its drivers will require much more data, a greater understanding of evolutionary pressures on deep-sea life histories, and better estimates of food availability. The potential explanatory power of metabolic rate in our data set, though limited by sample size, is a promising step in a fully predictive framework. Adding metabolic rate data to the top mixed model increased $R^2$ values from 47–61% when explaining $K$, and from 27–47% in explaining $K_{GDY}$. $R^2$ values for explaining AIM and $AIM_{GDY}$ also increased drastically when metabolic rate data was added (38–67% for AIM, 22–61% for $AIM_{GDY}$; B. splendens omitted). This would suggest that while our models are not yet fully predictive due to a limited data set, given sufficient environmental, phylogenetic, and metabolic rate data, a predictive framework for the growth of deep-sea demersal fishes is plausible.

Hydrostatic pressure is nearly perfectly correlated with depth and is a variable not directly explored in this analysis. Changes in pressure are known to af-
fect cell membrane fluidity, enzyme function, and rates of biochemical reactions in fish, and deep-sea fishes possess specific adaptations to tolerate high pressures, such as increased proportions of unsaturated fatty acids in cell membranes, pressure-adapted enzymes, and high concentrations of intracellular enzyme-stabilizing piezolytes (Somero 1990, Yancey & Siebenaller 2015, Gerringer et al. 2017). These adaptations, while necessary for tolerating high pressures, have been found to decrease enzyme efficiency and/or incur associated energetic costs. For example, fishes living below 500 m possess versions of dehydrogenases with reduced pressure sensitivity, but at the cost of catalytic efficiency (Somero 1990). Catalytic efficiency can be increased by simply producing higher concentrations of enzyme, but this increased production requires energetic input. As growth rate is the summed product of rates of many metabolic reactions, some of the explanatory power of depth we observed in predicting growth rates may arise mechanistically from increasing energetic costs associated with maintaining enzyme function and catalytic efficiency with increasing pressure at depth.

It is also worth noting that this work focused on demersal fishes, a group on which most marine growth studies have focused due to their wide commercial exploitation relative to deep pelagic species. In the pelagic, there is less research relating growth rates and depth, though Childress et al. (1980) found that *K* values generally followed a ‘slower, deeper’ trend. However, bathypelagic fishes grew faster than shallower mesopelagics in terms of yearly increases in caloric content. That study compared only 10 species and estimated age with whole otoliths, which have been shown to underestimate age particularly in deep-sea species, potentially obfuscating patterns of growth (Campana 2001). However, many of the species analyzed by Childress et al. (1980) have small otoliths, which makes more accurate aging methods such as transverse sectioning less feasible. Regardless, further growth studies on open-ocean fishes are necessary to determine whether they follow similar rates of decline in growth rate with depth as demersal fishes.

A number of past studies could not be included here due to procedural issues and lack of age-validation approaches, which hampered the present analysis. For instance, further work is needed to untangle some of the covariation in depth and temperature by collecting a data set with as many species as possible from the Mediterranean region, where temperature remains constant and relatively warm with depth. Twenty sources of age and size data from Mediterranean populations were examined, but 75% used whole otoliths without validation for age estimation and were rejected for quality control reasons (see ‘Supplementary Methods’ in Supplement 3). It is possible that annuli counts from viewing whole otoliths indeed accurately reflect the age for some or even all of these species, but without radiometric or other independent validation of the ring periodicity, the risk of underestimating age and overestimating growth rate was concern enough to exclude these studies (Campana 2001). More studies of otolith sections or using age validation are needed from warm deep-sea regions. We also acknowledge that the shallow-water growth literature is much broader than synthesized here. We attempted to include as much data as possible that fit quality controls and we focused primarily on including data from phylogenetic families that had representatives from across the depth gradient.

Determining the drivers of growth across demersal fishes is one component of understanding their resiliency to anthropogenic impacts, which continue to expand in the deep sea. Historically, studies of the life histories of deep-sea fish focused on direct exploitation, with some evidence that faster-growing stocks are more resilient to fishing mortality, though this trend is highly dependent on how ‘fishing mortality’ is defined (Koslow et al. 2000, Bailey et al. 2009, Juan-Jordá et al. 2015, Pinsky & Byler 2015, Färber et al. 2020). Past age and growth studies have focused on the upper bathyal (<1500 m), where commercially harvested species live. However, future threats to the ocean from deep-sea mining (reviewed in Glover & Smith 2003) and climate change (Sweetman et al. 2017) will occur at greater depths and present a real need for more information. For example, deep-sea mining is likely to commence at an industrial scale within the next few decades, with potentially decadal-scale disruptions to both benthic and pelagic habitats (Stratmann et al. 2018, reviewed in Glover & Smith 2003). Though the spatial and temporal extent of disruption is not yet fully clear, at current proposed depths of operation on seamounts and/or abyssal plains, it would certainly affect some of the deepest, slowest growing species in our analyses (*C. armatus, Antimora rostrata, H. atlanticus*, etc.). In terms of broad-scale anthropogenic climate change, recent estimates suggest abyssal ocean temperature may increase by 1°C by the end of the century, with concurrent decreases in the flux of POC (Sweetman et al. 2017). Deep-dwelling Antarctic fishes have no cooler refuge from warming waters, and may be at highest
risk. As such, we require more growth and life history data from the lower bathyal, abyssal, and hadal zones to fully understand drivers of growth across the full depth gradient and inform assessments of susceptibility to changing habitat conditions.

Deep-sea demersal fishes fulfill critical ecological roles as mid-to-top level predators and may be responsible for direct delivery of carbon from the mesopelagic to the seafloor via consumption of vertical migrants (Trueman et al. 2014, Drazen & Sutton 2017). These fishes may represent an important component of biogeochemical cycling through deep carbon deposition and burial. It is therefore essential to establish a fully predictive framework with which to gauge the susceptibility of these fishes to exploitation and future human stressors and to conserve their roles in both deep-sea ecosystems and the marine carbon cycle.

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