The life history of the invasive mullet, *Osteomugil engeli* (Bleeker, 1858) in Hawaiian estuaries

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Abstract Although mullets (*Mugilids*) are important estuarine fishes across the world, the biology of many species of mullet have yet to be determined. This is of concern for mullets that are introduced species with unknown impacts on native ecosystems. Here we assess the life history of *Osteomugil engeli*, an introduced species in Hawai‘i. Two populations (Hilo Bay estuary, Hawai‘i Island and Maunalua Bay estuary, Oahu) were compared to determine population demographics, spawning season, and size and age at maturity. There were no differences observed between life history metrics between populations sampled. *O. engeli* had year round spawning with recruits and adults co-occurring in similar habitats throughout the year. There was no difference in size (L50) and age (A50) at sexual maturity for females and males (L50: \( \chi^2 = 0.39, p\text{ value} = 0.53; A50: \chi^2 = 0.29, p\text{ value} = 0.59\)), with *O. engeli* reaching sexual maturity at 140.1 mm (CI: 139.5–141.6 mm) and 208 days (CI: 204.2–220.9 days). We observed a 4:1 female to male sex ratio and females were found to be on average 23 mm larger than males (t = 7.19, p value<0.01). *O. engeli* life history is compared to native species in Hawai‘i and among other *Mugilids* that have been introduced elsewhere. Life history comparisons suggest that introduced mullet populations are better able to respond and adapt to changing habitat and environmental conditions than the native mullet.

Keywords Mullet · Invasive species · Life history · Estuaries

Introduction

*Mugilids* are a ubiquitous family found in coastal waters worldwide that can dominate fish fauna in favorable habitats. Dominance of most grey mullet species, especially in estuaries, is directly attributable to their distinct position at the base of coastal food webs as large-bodied detritivores. Estuarine sediments are higher in particulate organic matter and microphytobenthos production compared to marine sediments making estuaries the preferred habitats of sedimentivory mullets, such as *Mugil cephalus*. Mullets are also economically important fish both commercially and recreationally as well as being vital for subsistence fishing in the Pacific Islands, Asia, Africa, and elsewhere because these fish convert primary productivity directly into fisheries production (Nishimoto et al. 2007; Whitfield et al. 2012; Sakihara et al. 2017).
The ability to forage primary for microphytobenthos and small filamentous algae may also, in part, contribute to the success of grey mullets as invasive species. As a result of human activities three mullet species are reported to have established populations outside of their native ranges (Randall 1987; Matishov and Luzhnyak 2007; Zenetos et al. 2010; Halim and Rizkalla 2011). A native to the Indian Ocean, the keeled mullet Liza carinata colonized the South-eastern Mediterranean as a Lessepsian immigrant species (Roll et al. 2007). This small-bodied mullet has limited fisheries value in its native range (Torcu and Mater 2000); however, along the Egyptian coast (Suez Bay) where this species is invasive, 18% of the total catch from 2002 to 2012 was comprised of L. carinata (Hefny et al. 2016). Planiliza haematocheila (also known as Mugil so-iuy or Liza haematocheila), a valuable fisheries and aquaculture species throughout its native range of the western North Pacific and the Sea of Japan, was intentionally introduced to the Azov and Black Seas in the late 1970s and early 1980s in efforts to improve aquaculture in the region (Okumus and Basçınar 1997). Osteomugil engeli (also known as Mugil engeli, Valamugil engeli, Moolgarda engeli, Chelon engeli, Agonostoma birarue, or Mugil kandavensis) has expansive native distribution in Oceania, which includes Indonesia, Melanesia, Micronesia, and Polynesia as well as warm waters of Northwestern Australia. Interestingly, O. engeli did not establish in the geographically isolated Hawaiian archipelago until it was accidentally included in intentional introduction of a baitfish, Marquesan sardine Sardinella marquesensis, during the 1950s (Randall 1987). O. engeli do not substantially contribute to any fisheries in their native range (Indo-Pacific ranging from East Africa to the Marquesan and Tuamoto islands; Froese and Pauly 2018) or as exotic species in Hawai‘i (Sakihara et al. 2017). This is likely as a result of its small size (up to 224 mm FL; Peyton et al. 2015) and restrictions on the mesh size of cast nets commonly used to harvest mullet in Hawai‘i. Additionally, O. engeli can be confused with similar looking striped mullet, M. cephalus (appearing as juveniles of this species), therefore often released by pole and line fishers as undersized M. cephalus (known locally as ama ama).

The Hawaiian Islands have two native mullet species, Mugil cephalus (Linnaeus, 1758) and Neomyxus leuciscus (Gunther, 1872). All three species co-occur in estuary habitats in Hawai‘i (Nishimoto et al. 2007). Both O. engeli and N. leuciscus utilize nearshore marine waters and estuaries during its adult life, whereas M. cephalus is more reliant on estuary and brackish environments (Nishimoto et al. 2007; Ghasemzadeh 2015). Since introduction, O. engeli became well established with reports of up to 80% of all mullet abundance comprised of O. engeli in Kaua‘i estuaries in the 1970s (Eldredge 1994; Nishimoto et al. 2007). They currently remain a dominant species in Hawaiian estuaries and are known to occur across the main Hawaiian Islands (Nishimoto et al. 2007).

Understanding introduced species life history is important for risk assessments (Kolar and Lodge 2002) and management of invasive species (Morris et al. 2009). Life history metrics allow for the assessment of population growth rates and comparisons to native species to better understand potential species competition (Holt 1977). For example, overlap in spawning or recruitment periods could lead to competitive interactions for resources or other potential indirect impacts such as aggregations of spawners or recruits resulting in increased attention of predators.

Additionally, environment has been found to influence reproductive characteristics (Caselle et al. 2011; Taylor and Choat 2014; Schemmel and Friedlander 2017) and growth in fishes (Gust et al. 2002; Caselle et al. 2011; Donovan et al. 2013). Therefore, it is important to conduct life history studies for introduced species in their introduced range. Furthermore, the sample design should also consider small-scale environmental plasticity to capture the variability within the region.

The biology and life history of O. engeli has not been studied in its native range or introduced range, impacting our ability to understand their potential impacts on native populations and estuaries as well as to develop management plans for this invasive species in Hawai‘i. We examined age, growth, and reproduction of this abundant invasive in Hawai‘i from two estuaries where O. engeli are known to be abundant: Hilo Bay and Maunalua Bay. Our objectives were to determine the life history of O. engeli in its recipient range for both populations including: population size structure, age and growth, female and male size (L50) and age (A50) at sexual maturity, and spawning season.
Methods

Sampling locations and design

*Osteomugil engeli* population characteristics were compared in two estuaries: Maunalua Bay (Wailupe) on Oahu and Hilo Bay (Reeds Bay) on Hawai‘i Island (Fig. 1). Maunalua Bay is located on the southeastern shore of Oahu and is an 8.7 km long embayment estuary with freshwater inflow from both perennial and intermittent streams (surface water) as well as coastal ground water. One of 10 watersheds that drain into Maunalua Bay, Wailupe watershed is located in the center of the bay and receives intermittent surface water input, although most of its freshwater is from coastal ground water discharge. At the sampling station substratum is composed of ca. 70% sediment and ca. 30% carbonate with a depth range to 0.6 m. Hilo Bay, a crescent shaped 6.5 km long embayment estuary, is located on Hawai‘i Island’s eastern shoreline. It is protected by a 4.82 km long man-made break-wall with freshwater influxes derived from two perennial rivers (Wailuku and Wailoa), smaller perennial streams, and numerous subterranean springs. The sampling station at Reeds Bay (although locally known as a bay, it is a cove within Hilo Bay) is located at the eastern shore of Hilo Bay with freshwater inflow primarily coming from coastal ground water discharge and to a lesser extent through intermittent surface water. The substratum at Reeds Bay is composed of ca. 50% sediment and 50% basalt with a depth range to 1 m.

Maunalua Bay sampling occurred from September 2015 to August 2017 and Hilo Bay from July 2016 to August 2017. A fixed sampling station (100 m) at each site was selected based on data from earlier surveys (DLNR-DAR unpublished data) showing that, within each estuary sampled, *O. engeli* utilizes this area as a sub-habitat hotspot. Sub-habitat hotspot is defined as fixed linear distances along a shoreline that encompass areas that have consistently yielded fish captures from random net casts (monofilament cast nets, see below) with >75% probability of capture. The linear boundaries for each sampling station (sub-habitat hotspot) were Maunalua Bay N21.27611 and W-157.76176 to N21.27608 and W-157.76180, and Hilo Bay N19.72559 and W-155.06256 to N19.72649 and W-155.06288).

Monofilament cast nets (2.44 m diameter with 0.635 cm mesh size) were used to sample fish for two different sampling methods. Visual (non-random) method was completed first, followed by a 15-min rest interval that allowed fish to recover from disturbance and

Fig. 1  Sampling locations for *Osteomugil engeli* in Hilo Bay and Maunalua Bay estuaries
then, if needed, the area was sampled using Probability of Encounter (POE, systematic) approach. Since Miranda (2007) recommends that length-frequency histograms with 1-cm intervals require \( \geq 375 \) fish to estimate within 10% precision with 80% confidence, whenever possible a minimum of 300 individuals were sampled per month and \( O. \text{engeli} \) population data were pooled into 3-month seasons. When <300 individuals were measured per month using visual sampling then POE method was used to supplement the sample effort or, as frequently occurred, to corroborate that \( O. \text{engeli} \) had occurred in low abundance at the time of sampling (Table 1). All sampling took place in the mornings on haphazardly selected dates.

During visual sampling, a fish surveyor actively sampled for \( O. \text{engeli} \) by locating schooling fish within the selected sampling station boundaries and casting on the first school observed. Additionally, a fish surveyor had 32 min of total searching time, which was tracked by an observer on shore using a stopwatch. Transferring fish samples from net to bucket, measurements of specimens, and reloading of a cast net were not counted against searching time. On occasions when no \( O. \text{engeli} \) were visually located or when visibility was limited, haphazard casts were made in microhabitats until all replicates were taken. Some of these microhabitats included coastal ground water plumes, known to attract schools of \( O. \text{engeli} \) (unpublished DLNR-DAR data). Sixteen casts were taken per sampling event.

POE method cast net sampling began at a sampling station boundary with each subsequent cast taken at a distance of 20 paces away from the previous cast (ca. 12 m) until the surveyor reached the other boundary. At this boundary, the fish surveyor turned around and continued sampling until reaching another boundary or until all samples were taken (\( n = 16 \)). Resampling too close to a position previously sampled along the shoreline was avoided by the fish surveyor offsetting his paces by subtracting five paces on the first cast after the turn was made and then resuming the 20 paces to take each subsequent cast. For both methods, fish are held in aerated buckets, measured to the nearest mm, and euthanized by immersion ice bath using a portable cooler in the field. Transferring fish samples from net to bucket, measurements of specimens, and reloading of a cast net were not counted against searching time. On occasions when no \( O. \text{engeli} \) were visually located or when visibility was limited, haphazard casts were made in microhabitats until all replicates were taken. Some of these microhabitats included coastal ground water plumes, known to attract schools of \( O. \text{engeli} \) (unpublished DLNR-DAR data). Sixteen casts were taken per sampling event.

To determine female and male size \((L_{50})\) and age \((A_{50})\) at sexual maturity and spawning season, \( Osteomugil \text{engeli} \) were sampled monthly from both Maunalua Bay (September 2015 – December 2016) and Hilo Bay (December 2015 – December 2016). Approximately 20 individuals with \( \geq 110 \) mm fork length (FL) were collected each month at both study sites using monofilament cast nets (2.44 m diameter with 0.635 cm mesh size). A larger area was used to sample fish at both sites for these analyses as compared to sampling populations (see above). This helped to broaden collection of specimens that spanned size ranges of sub-adult to adult and to minimize impacts of extracting fish from the designated sampling areas. Maunalua Bay specimens were collected from the Wailupe watershed N21.27623 and W-157.76095 to N21.27232 and W-157.77069. Hilo Bay samples were collected from Reeds Bay N19.72559 and W-155.06256 to N19.72649 and W-155.06288. Sampling dates were the same as dates as the population study (see above).

Fish surveyors targeted observed schools of sub-adult to adult sized \( O. \text{engeli} \) for sampling. Fish were held in aerated buckets, measured to the nearest mm, and euthanized by immersion ice bath using a portable cooler in the field with any excess fish returned to the estuary. For each sampling date, the goal was to collect fish from the size range present on that day by sampling the first two fish caught for each 10 mm FL size bin (110 to 200 mm FL). Larger sizes (\( \geq 190 \) mm FL) were less commonly sampled throughout the study, therefore, smaller sizes had up to three replicates for each FL size bin to sample approximately 20 individuals per month at each site.

To determine age and growth of \( O. \text{engeli} \) populations, up to five fish per 10 mm FL size bin \((\geq 20 \) mm FL) were sampled opportunistically at both study sites on haphazardly determined dates from December 2015 – December 2016 using monofilament cast nets (2.44 m diameter with 0.635 cm mesh size) for individuals \( \geq 28 \) mm FL and dip nets (0.15 cm mesh size) for \( \leq 27 \) mm FL. Dip nets were used for individuals \( \leq 27 \) mm FL because at this size the fish are too small to be consistently sampled with cast nets used in this study. Fish were euthanized by immersion ice bath using a portable cooler in the field.

Samples were transported to the DLNR-DAR Wailoa Fisheries Research Station (Hawai‘i) and DLNR-DAR Ānuenue Fisheries Research Center (Oahu) where metrics were recorded (fork length (mm); total fish weight (MF, 0.00 g)) and otoliths were extracted. Sagittal otoliths were removed, rinsed in water, and stored until further processing. Fish were dissected to extract gonads, gonad weight (MG, 0.00 g), examined for macroscopic gender identification and a midsection of one
Table 1  Sampling effort (n = number of casts) and population (N) sampled using visual method and probability of encounter (POE) for populations of *Osteomugil engeli* at Wailupe (Maunalua, Oahu) and Reeds Bay (Hilo, Hawai‘i) by sampling date each month.

| Site       | Date       | Visual method | POE | Visual method | POE | Total N |
|------------|------------|---------------|-----|---------------|-----|---------|
| Wailupe    | 22 Sep 2015| 16            | 16  | 271           | 0   | 271     |
|            | 26 Oct 2015| 16            | 0   | 376           | –   | 376     |
|            | 23 Nov 2015| 16            | 0   | 692           | –   | 692     |
|            | Dec 2015    | 16            | 16  | 50            | 46  | 96      |
|            | Jan 2016    | 16            | 16  | 197           | 27  | 224     |
|            | Feb 2016    | 16            | 16  | 58            | 3   | 61      |
|            | 17 Mar 2016 | 16            | 16  | 208           | 0   | 208     |
|            | 25 Apr 2016 | 16            | 16  | 186           | 28  | 214     |
|            | 16 May 2016 | 16            | 16  | 249           | 55  | 304     |
|            | 29 Jun 2016 | 16            | 16  | 267           | 1   | 268     |
|            | 7 Jul 2016  | 16            | 16  | 249           | 64  | 313     |
|            | Aug 2016    | 16            | 0   | 815           | –   | 815     |
|            | 2 Sep 2016  | 16            | 0   | 539           | –   | 539     |
|            | 20 Oct 2016 | 16            | 16  | 141           | 2   | 143     |
|            | 22 Nov 2016 | 16            | 16  | 75            | 0   | 75      |
|            | 13 Dec 2016 | 16            | 16  | 122           | 39  | 161     |
|            | 19 Jan 2017 | 16            | 16  | 73            | 89  | 162     |
|            | 24 Feb 2017 | 16            | 16  | 246           | 72  | 318     |
|            | 30 Mar 2017 | 16            | 16  | 245           | 23  | 268     |
|            | 19 Apr 2017 | 16            | 16  | 95            | 24  | 119     |
|            | 17 May 2017 | 16            | 16  | 328           | 91  | 419     |
|            | 28 Jun 2017 | 16            | 0   | 402           | –   | 402     |
|            | 28 Jul 2017 | 16            | 0   | 615           | –   | 615     |
|            | 17 Aug 2017 | 16            | 16  | 17            | 8   | 25      |
| Reeds Bay  | 9 Dec 2015  | 16            | 16  | 145           | 0   | 145     |
|            | 12 Jan 2016 | 16            | 16  | 227           | 0   | 227     |
|            | 11 Feb 2016 | 16            | 16  | 38            | 1   | 39      |
|            | 3 Mar 2016  | 16            | 16  | 52            | 58  | 110     |
|            | 19 Apr 2016 | 16            | 0   | 0             | –   | 1       |
|            | 19 May 2016 | 16            | 16  | 301           | 0   | 301     |
|            | 2 Jun 2016  | 16            | 16  | 111           | 0   | 111     |
|            | 7 July 2016 | 2             | 0   | 319           | –   | 319     |
|            | 16 Aug 2016 | 5             | 0   | 305           | –   | 305     |
|            | 15 Sep 2016 | 7             | 0   | 308           | –   | 308     |
|            | 12 Oct 2016 | 16            | 0   | 418           | –   | 418     |
|            | 22 Nov 2016 | 16            | 16  | 80            | 1   | 81      |
|            | 14 Dec 2016 | 16            | 16  | 1             | 3   | 4       |
|            | 4 Jan 2017  | 16            | 0   | 585           | –   | 585     |
|            | 15 Feb 2017 | 16            | 0   | 637           | –   | 637     |
|            | 8 Mar 2017  | 16            | 0   | 286           | –   | 286     |
|            | 12 Apr 2017 | 16            | 0   | 983           | –   | 983     |
|            | 10 May 2017 | 16            | 0   | 613           | –   | 613     |
|            | 5 Jun 2017  | 16            | 0   | 1027          | –   | 1027    |
|            | 6 Jul 2017  | 16            | 0   | 938           | –   | 938     |
|            | 4 Aug 2017  | 16            | 0   | 1504          | –   | 1504    |
gonad lobe was excised and stored in buffered formalin for a minimum of three days.

Tissue processing and histological assessment

Gonads were rinsed overnight in fresh water and rinsed with a series of alcohol dilutions (30% and 50%) for 30 min and stored in 70% ETOH until processing. Gonad samples were sent to the University of Hawai‘i John A. Burns School of Medicine histopathology lab, embedded in paraffin, sectioned at 5 μm, and counter stained with Hematoxylin and Eosin.

Female reproductive state was assigned based on a schema modified from Brown (Brown-Peterson et al. 2011) (Table 2; Fig. 2). Resting females were distinguished from immature females by evidence of past spawning events including a thick ovarian wall, post-ovulatory follicles, and atresia (Fig. 2). Spawning capable females were distinguished from developing females by the presence of late stage vitellogenic oocytes (VTIII) and actively spawning females were identified by the presence of hydrated oocytes (Fig. 2). Males were classified as mature if spermatozoa were present (Fig. 3).

Student T-Tests were done to compare male and female fish sizes (FL) at and between locations. Difference in female and male size between locations was assessed with a one-way ANOVA with Tukey Honest Significant Difference Test.

Sexual maturity and spawning seasons

Size at sexual maturity (L50) and age (A50) at sexual maturity was assessed as the size or age at which 50% of individuals of a given gender are reproductively mature based on a logistic regression model with binomial family and logit link function (Chen and Paloheimo 1994). Size at maturity was assessed using 5 mm fork length bins and age at maturity was assessed using 10 day age bins. Likelihood ratio tests were done to assess differences L50 between locations and between genders at each location.

Spawning season was assessed at Hilo Bay and Maunalua Bay by comparing monthly gonadosomatic index (GSI). GSI was assessed as:

\[
GSI = \frac{GM}{GFM} \times 100
\]

where GM is the gonad mass and GFM is the gonad free fish mass.

Spawning season was also confirmed through analysis of the monthly frequency of reproductive states. Months with females in either spawning capable and actively spawning status were considered within the spawning season.

Age determinations

The right sagittal otolith was analyzed first (92% if the time) unless it was missing or broken, in which case the left sagittal otolith was used. Otolith length (mm) was measured from the anterior to posterior end with a caliper and otolith weight (g) was measured using a sensitive scale. Each otolith was mounted on a glass slide using Crystal bond™ with the anterior end protruding over the slide, such that the visible opaque otolith core was inside of the edge of the slide. The anterior end was then ground against 40 μm grit lapping film using a Crystal master™ until it was flush with the slide. This polished end is then mounted face down on the slide and the posterior end is

### Table 2 Reproductive state classification schema for female Osteomugil engeli

| Reproductive state      | Mature | Diagnostics                                                                                       |
|-------------------------|--------|--------------------------------------------------------------------------------------------------|
| Undeveloped             | No     | Ovaries with oogonia and primary growth (PG) oocytes (chromatin-nucleolus and perinucleolar) present. Thin ovary wall. |
| Developing              | Yes    | Cortical alveolar oocytes, and/or early vitellogenic oocytes, VTI and/or VTII, present.          |
| Spawning Capable        | Yes    | Presence of late stage vitellogenic oocytes (VT III), identified by dramatic increases in oocyte size and uniform distribution of yolk. |
| Actively Spawning       | Yes    | Contains fully hydrated oocytes. May contain postovulatory follicles (POFs) and retained oocytes post spawning. |
| Resting                 | Yes    | Regressing or regenerating ovary. The ovary wall is thick, does not contain healthy vitellogenic oocytes, and the ovary may contain unabsorbed material from past spawning events (atretic oocytes and POFs). |
ground down to produce a transverse section. Series of manual polishing with $9 \, \mu m$ and $0.3 \, \mu m$ lapping films are used until daily growth increments are visible under a compound microscope.

Consecutive images are taken via a camera mounted on the compound microscope at 200x (eye piece at 10x and objective at 20x) magnification (otolith images) (Fig. 4). Growth increments were counted, and replicate counts were completed for a subset of the otoliths. An otolith was rejected if the counts deviated >10%, otherwise the mean count was used as the age.

Age and growth

To determine growth rate, three different models were compared: von Bertalanffy, linear, and power. The von Bertalanffy growth (Ricker 1987; Choat and Robertson 2002) was assessed as

$$L_t = L_\infty \left[1 - e^{-k(t-t_0)}\right]$$

where, $L_t$ is length at age $t$, $L_\infty$ is mean maximum asymptotic length, $t$ is age in days, $t_0$ is the theoretical age at which length is 0, and $k$ is the Brody growth

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**Fig. 2** Female *Osteomugil engeli* reproductive stages. a) Immature ovary with thin ovary wall and primary growth oocytes b) developing ovary with primary growth and vitellogenic I oocytes, c) spawning capable ovary with vitellogenic III oocytes, d) actively spawning individual with hydrated oocytes, e) resting individual with evidence of recent spawning including post ovulatory follicles, f) resting ovary with thick ovarian wall, primary growth oocytes, and large melano-macrophage centers. Ovarian wall (OW), primary growth oocytes (PG), vitellogenic I (VTI), vitellogenic III (VTIII), hydrated (H), post ovulatory follicle (POF), melano-macrophage center (MMC). Scale bar is 100 \( \mu m \)

**Fig. 3** Male *Osteomugil engeli* reproductive states a) Immature male with spermatogenic tissue. b) Mature male with spermatozoa. Spermatogenic tissue (ST), spermatozoa (S). Scale bar is 100 \( \mu m \)
coefficient. Best model was selected based on coefficient of determination ($r^2$) and model differences in Akaike Information Criterion ($\Delta$AIC) (Akaike 1973).

Mean generational turnover (GT) is an estimate of the average time for a new generation to replace the last generation assuming a stable population (Gaillard et al. 2005). Generational turnover was estimated with the following equation (Gaillard et al. 2005; Depczynski and Bellwood 2006).

$$GT = AM + [(T_{max} - AM)/2]$$

where AM is the age at first female maturation ($L_{50}$) and $T_{max}$ is the maximum age. All statistical tests were performed in R using packages stats, car, and boot (R Core Team 2016).

**Results**

**Population demographics**

Over 16,000 *Osteomugil engeli* individuals were monitored using Probability of Encounter (POE) and visual methods with 7088 individuals from Hilo Bay and 8942 from Maunalua Bay. Catch rates fluctuated greatly with an average monthly catch from Hilo Bay of $425 \pm 401$ ($\mu \pm sd$) and Maunalua Bay $295 \pm 203$ (Table 1). *O. engeli* observed ranged...
Fig. 6  Length-frequency distribution of *Osteomugil engeli* sampled from Wailupe, Maunalua Bay, Oahu from fall 2015 through summer 2017. a) Fall, b) winter, c) spring, and D) summer

Fig. 7  Length frequency distribution of *Osteomugil engeli* collected using cast nets and dip nets from a) Hilo Bay (Hawai‘i) and b) Maunalua Bay (Oahu)
in size from 20 to 229 mm. Recruitment size classes (<40 mm) through adult size classes (>140 mm) were encountered each month, co-occurring in the same habitat. Seasonal abundance of each size class was highly variable. Further, while there was no clear temporal pattern in recruitment, recruitment was detected year-round (Figs. 5 and 6).

*O. engeli* biologically sampled ranged from 30 mm to 209 mm (FL). Female and male size distributions did not differ between locations ($\chi^2 = 0.37$, $p$ value = 0.54). Female *O. engeli* were found to be on average 23 mm larger than males ($t = 7.19$, $p$ value <0.01) (Fig. 7). We found a 4:1 female to male sex ratio from *O. engeli* sampled from a 100 mm–209 mm size distribution. There were 236 females and 84 males sampled from Maunalua Bay and 254 females and 37 males sampled from Hilo Bay.

Sexual maturity and spawning season

A total of 611 sampled *Osteomugil engeli* were included in the reproductive assessment. Female *O. engeli* size at maturity ($L_{50}$) and age at maturity ($A_{50}$) did not differ between locations ($L_{50}$: $\chi^2 = 0.42$, $p$ value = 0.51; $A_{50}$: $\chi^2 = 0.04$, $p$ value = 0.85). There was no difference in size ($L_{50}$) and age ($A_{50}$) at sexual maturity for females and males ($L_{50}$: $\chi^2 = 0.39$, $p$ value =0.53; $A_{50}$: $\chi^2 = 0.29$, $p$ value = 0.59). *O. engeli* reach sexual maturity at 140.1 mm (CI: 139.5–141.6 mm) and 208 days (CI: 204.2–220.9 days) (Fig. 8).

*O. engeli* exhibited group synchronous sensu oocyte development, with a synchronous batch of oocytes developing for each spawning event (Wallace and Selman 1981). *O. engeli* exhibited year-round spawning. High gonadosomatic values and spawning capable females were found throughout the year in both Hilo Bay and

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**Fig. 8 Osteomugil engeli size and age at maturity using 5 mm fork length bins and 10 day age bins. a) $L_{50} = 140.1$ mm (CI: 139.5–141.6 mm) and b) $A_{50} = 208$ days (CI: 204.2–220.9 days). No difference in size or age at maturity was found between females and males.**
Maunalua Bay (Figs. 9 and 10). Actively spawning females were found from April through December. Year-round spawning activity of this species was also reflected in year-round recruitment to estuaries in Hawai‘i (Figs. 5 and 6). There were, however, large fluctuations in young-of-year observed that had no detectable pattern spatially or temporally (Figs. 5 and 6).

Age and growth

A total of 227 Osteomugil engeli were aged (81 from Hilo Bay and 146 from Maunalua Bay) spanning a size range from 30 mm to 209 mm. Ages ranged from 44 to 349 days ($\mu = 249.6$ days ±72.7 sd) for Hilo Bay and 36 to 429 days ($\mu = 221.7$ days ±81.2 sd) for Maunalua Bay. The oldest fish was 429 days old ($T_{\text{max}}$) and 185 mm FL. Generational turnover was estimated at 215 days or approximately seven months.

O. engeli, individuals with unknown gender, males, and females were assessed together. The von Bertalanffy growth model ($r^2 = 0.82$) was the best-fit model followed by power ($\Delta AIC = 21.0$, $r^2 = 0.81$) and linear ($\Delta AIC = 38.8$, $r^2 = 0.79$) growth models. O. engeli show asymptotic growth with growth rates slowing as fish reach maximum size ($FL = 278.71(1-e^{-0.0036(Age-19.7)})$, Fig. 11). Females and males exhibited differential growth patterns ($F = 2.69$, $p$ value<0.01), with males reaching smaller maximum sizes than females of a similar age (Fig. 11).

Discussion

Like many other smaller bodied mullet species, the life history of Osteomugil engeli was previously unknown. Hawai‘i populations of O. engeli were found to exhibit year round spawning and young size at maturity; both aspects of this species' life history that likely contribute to this its successful invasion of Hawaiian estuaries after an accidental introduction.

There were no observed differences in population growth rates, size at maturity, or spawning seasons between our two sampling locations. It is likely that different environmental conditions influence growth and reproductive characteristics between these two populations but since there was not a distinct spawning season we did not detect variable in the spawning season between locations. Additionally, due to a young size at
maturity, there was not much time for environmental influence on maturation. However, there were variable peaks in recruitment over the two years of monthly sampling, indicating that these recruitment peaks maybe in response to environmental factors yet to be determined.

We saw no size-dependent change in habitat use among juveniles, subadults, and adults because all were co-occurring. *O. engeli* (adults >140 mm) may be less abundant in our samples for several possible reasons. Adults likely have a larger home range than juveniles and our 100 m sampling area many have been too small to adequately sample full adult range. Although adults were observed to swim in schools through both sampling sites, compared to juveniles and subadults, adults were better at avoiding net capture by swimming in smaller, looser schools and rapidly moving to outside the sampling area.

Native *M. cephalus* spawn during the winter and spring in Hawai’i. *M. cephalus* spawning generally occurs from December – February (Honebrink 1990) but can fluctuate slightly from year to year (Schemmel unpublished research). Annual recruitment of *M. cephalus* to estuary habitats occurs in spring (Major 1978), whereas we found recruitment of *O. engeli* year-round.

The introduced populations of *O. engeli* in Hawai’i become reproductively mature at a young age (208 days) and small size (140 mm) compared to native mullet. *M. cephalus* size and age at sexual maturity is

Fig. 10 Reproductive state frequency of *Osteomugil engeli* in a) Hilo and b) Maunalua Bay. Refer to Table 2 for descriptions of reproductive state. AS = Actively Spawning, SC = Spawning Capable, D = Developing, R = Resting, U=Undeveloped
unknown for Hawai’i populations however studies from other location suggest that *M. cephalus* reach sexual maturity around 3 to 4 years old (Tung 1981; Smith and Deguara 2002; Whitfield et al. 2012) and at a size around 310–380 mm SL (Apekin and Vilenskaya 1979; Smith and Deguara 2002).

The life history of the other native mullet to Hawai’i, *Neomyxus leuciscus* (known locally as Uouoa), is not known. However, given the strong correlation between maximum length and size at maturity in tropical fishes (DeMartini and Howard 2016) and similar maximum sizes of *O. engeli* and *N. leuciscus* for Hawaiian populations (*O. engeli: 224 mm, N. leuciscus: 229 mm; Peyton et al. 2015*), it is likely that *N. leuciscus* reaches maturity at a similar size or age as *O. engeli* for Hawaiian populations. However, growth rate as well as gonadal maturation and spawning periods have not been reported for any *Neomyxus* species.

*Osteomugil engeli* life history was compared to other introduced mullet species (*Planiliza haematocheila* and *Liza carinata*) (Table 3). All three introduced mullet species spend at least part of their life history in estuaries and are microphagous species. *P. haematocheila* is an euryhaline mullet species that was purposely introduced to Black and Azov Seas because native mullets were in decline due to changing environmental conditions resulting from decreased freshwater discharges and overfishing (Starushenko and Kazanskii 1996). *L. carinata* is a small-bodied mullet similar in size to *O. engeli* that was introduced in the South Eastern Mediterranean. *P. haematocheila* and *L. carinata* were
found to be seasonal spawners, May to July, and August to October, respectively (Okumuğ and Başçınar 1997; Torcu and Mater 2000). Like M. cephalus, P. haematocheila is a larger bodied mullet (up to 667 mm TL) with an age and size at maturity estimated at 430 mm and 4–5 years for females and 375 mm and 2–4 years for males (Okumus and Bascinar 1997; Torcu and Mater 2000). L. carinata has a similar size range as O. engeli and similar size at maturity of 125 mm for male and 143 mm for females (Hefny et al. 2016).

Management considerations

It is currently unknown if Osteomugil engeli impacts other estuarine fishes, especially native mullet species. Native and introduced mullet feed mainly on benthic diatoms (microphytophenthos) (Platell et al. 2006; Sakihara et al. 2017). However, native mullet (M. cephalus) have been found to prefer Melosira moniliformis, which is maintained by flood events of freshwater into marine environments (Julius 2007). Fishers even utilize M. moniliformis as bait for native mullet (Julius 2007). Intraspecific competition for microalgae in mullet has been identified for other mullet species (Mullus surmuletus) and found to effect foraging behavior, movements, and energy budget for the species (Levi and Francour 2004; Longepierre et al. 2005). The modification of streams and estuary habitats alters flow regimes and microalgal community structure and thus may give O. engeli an advantage over native mullet (Julius 2007). Also, invasive plant canopies that have overgrown many estuaries in Hawai‘i can have negative effects on M. cephalus with only limited impacts on the generalist O. engeli (Sakihara et al. 2017). Only the other hand, degradation of Hawaiian estuaries and/or overfishing of M. cephalus may allow the population of O. engeli to increase similar to the scenario for introduced populations of L. haematocheila in Azov and Black Seas (Okumus and Başçınar 1997).

M. engeli may affect other important fishery species that utilize estuaries for part of their life cycle. For example, estuaries are nurseries and important habitats for many of Hawai‘i’s prized fishery species including jacks (Hawaiian names: ulua, papio, omul; Caranx ignobilis, Caranx melampygus) (Smith and Parrish 2002). Mullet are an important diet component of juvenile jacks (Smith and Parrish 2002) and likely other piscivores and therefore may have a positive effect on those species.

Conclusions

The introduced mullet’s, Osteomugil engeli, life history shows that they have a population growth rate advantage over native mullet. O. engeli reproduce at smaller and younger sizes than either of the native mullet species and spawn year-round. The shorter generational turnover rates for the introduced mullet also allow populations to respond and adapt to changing conditions faster than the native mullet. Growth rates are faster and age at maturity earlier than native populations. Destruction and alteration of stream and estuary ecosystems may exacerbate impacts of O. engeli on native systems. Further research is needed to map the distribution of the introduced mullet in Hawai‘i and to better understand species and ecosystem impacts of this introduction.

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