Vegetation and macrozoobenthos diversity in the Percut Sei Tuan mangrove forest, North Sumatra, Indonesia

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Abstract. Hasibuan IM, Amelia R, Bimantara Y, Susetya IE, Susi, M, Basyuni M. 2021. Vegetation and macrozoobenthos diversity in the Percut Sei Tuan mangrove forest, North Sumatra, Indonesia. Biodiversitas 22: 5600-5606. The mangrove forest offers a habitat for a variety of marine species, including macrozoobenthos. Macrozoobenthos live at the bottom of the waters on hard to muddy substrates, where they break down mangrove debris, ensuring microorganisms' ease in decomposing organic matter into inorganic materials that provide nutrients to producers in the water. The study aimed to determine the vegetation and macrozoobenthos diversity in the Percut Sei Tuan mangrove forest, North Sumatra, Indonesia. This research was executed with vegetation analysis, employing a sampling method in the form of plots, as many as three transects and 66 plots. The sampling of macrozoobenthos was conducted on the substrate surface, roots, and mangrove trunk for three stations and nine plots with a size of 1 m x 1 m. We found 10 true mangrove species where Avicennia marina was the most abundant, found at each seedling, sapling, and tree-stage with an important value index (IVI) of 64.62, 75.73%, and 92.12%, respectively. Overall, the number of individuals in the species found was 1666.67/Ha at the seedling stage, 1090.91/Ha at the sapling stage, and 842.42/Ha at the tree stage. Seven macrozoobenthos species were found and classified into four classes of Gastropoda, Polychaeta, Mollusca, and Arthropoda, with population abundance values being 9.33 ind/m² at a station I, 7.44 ind/m² at station II, and 8.78 ind/m² at station III. The diversity index (H') of the observed location was determined to be 1.63-2.06, which was classified into the medium category, while H' macrozoobenthos ranging from 1.54-1.81 were classified into the low to moderate category. The observed parameters' physical and chemical factors presented that the condition of the waters in the research area is still considered suitable for supporting macrozoobenthos life, with dissolved oxygen being strongly correlated to macrozoobenthos diversity.

Keywords: Biodiversity, community assemblage, macrozoobenthic, natural mangrove, physicochemical characteristic

INTRODUCTION

Mangrove forests are located in coastal areas and salt marshes which are affected by tides and can also grow in tropical and subtropical coastal areas (Balke et al. 2013, 2015; Bryan-Brown et al. 2020). Ecologically, mangroves play a role in producing oxygen and absorbing carbon dioxide, climate change mitigation efforts, as a water nutrition provider, and as a habitat for various types of animals even though the area comprises merely of 0.5% of the world's coastal areas (Alongi 2014; Murdiyarso et al. 2015).

Macrozoobenthos are a group of aquatic organisms that have a crucial part in aquatic ecosystems; it can live on hard to muddy substrates by making holes, crawling, sticking, burying both at the bottom and surface of the waters, and live by sticking to mangrove trees (Tapilatu and Pelasula 2012; Basyuni et al. 2018; Chen et al. 2018). It is often used as a bioindicator to assess the status of water quality due to its relatively low movement; a change in water quality will have a significant impact on macrozoobenthos (Bayan et al. 2016; Susetya et al. 2018).

Based on its role and functions, macrozoobenthos may decompose mangrove waste, facilitating the process for microorganisms to convert organic matter into inorganic materials and provide nutrients to producers in the waters (Muhammad et al. 2017). Macrozoobenthos found in mangrove forest areas come from Class Mollusca, Polychaeta, Bivalvia, Arthropoda, and Malacostraca (Basyuni et al. 2018). Several factors affect the population of macrozoobenthos: physicochemical properties in the aquatic environment, the entry of oxygenated water changing water quality and heavy metal content in sediments and water depths, as well as biotic factors such as mangrove plants being a source of food jointly affects the community structure and distribution of macrozoobenthos (Obolewski 2011; Li et al. 2020).

The current mangrove ecosystem continues to be damaged due to changes in land use and land cover without regard to environmental aspects, as in the mangrove forests of North Sumatra, Percut Sei Tuan district (Bryan et al. 2013; Sasmito et al. 2019; Basyuni et al. 2021). Percut Sei Tuan is a coastal area located on the East Coast of North Sumatra. As a coastal area includes rivers with several
tributaries, the reduced area of mangroves and their damage can also affect the abundance of fauna—especially the macrozoobenthos in the area (Susetya et al. 2021). The aim of the present study was to determine the diversity of vegetation and macrozoobenthos in the mangroves of Percut Sei Tuan, North Sumatra, Indonesia.

MATERIALS AND METHODS

Study area

The research for this study was conducted from February to March 2021 in the mangrove forest of Tanjung Rejo Village, Percut Sei Tuan district, Deli Serdang Regency, North Sumatra, Indonesia. The sampling locations were divided into three different stations: Station I is located at 3°44′03.09″-3°44′06.23″ North Latitude and between 98°46′22.14″-98°46′19.72″ East Longitude, Station II at N 3°44′04.02″-3°44′06.98″ and between E 98°46′22.94″-98°46′20.49″, and Station III at N 3°44′06.57″-3°44′09.23″ and between E 98°46′24.26″-98°46′21.81″ (Figure 1). The mangrove ecosystem on the coast of the Tanjung Rejo village contains vegetation that is either distributed naturally or planted by the local community.

Vegetation analysis

The vegetation analysis executed at the mangrove forest had three sampling stations, within which three plots (triplicates) were determined to collect basic forest stand attributes. The data collection technique, performed following Pearson et al. (2005), gathered information from three transects of 100 m inland perpendicular to the coastline, on transect I, there was as many as 25 plots, transect 2 had 23 plots, and transect 3 had 18 plots (Figure 1). Each plot measured 10 m x 10 m for trees, 5 m x 5 m for saplings, and 2 m x 2 m for seedlings. The identification of the species was carried out using a mangrove identification book which is referred to in both Kitamura et al. (2003) and Noor et al. (2006).

Frequency (F), Density (A), and Dominance (D)

Species frequency is a value that provides the distribution of the species in the plot.

\[
\text{Frequency (F)} = \frac{\text{the number of plots there are types}}{\text{sum of all plots}} \tag{1}
\]

Relative Frequency (RF) = \frac{\text{Frequency of a type} \times 100\%}{\text{All kinds of frequency}} \tag{2}

Abundance is a way of determining the number of individuals in a plot. The density value is calculated by the formula:

\[
\text{Density (A)} = \frac{\text{Number of individuals of a species}}{\text{Sample plot area}} \tag{3}
\]

Relative Density (RA) = \frac{\text{Density of species} \times 100\%}{\text{Abundance of all species}} \tag{4}

Dominance specifies the dominance of the species contained in the plot, the value of which is obtained by calculating the base area of a species and dividing it by the area of the entire plot.

\[
\text{BA} = \frac{\text{basal area of species}}{\text{Sample plot area}} \tag{5}
\]

\[
\text{RBA} = \frac{\text{Dominance of species} \times 100\%}{\text{Dominance of all species}} \tag{6}
\]

Where, BA: Basal area; RBA: Relative basal area.

Figure 1. Map of sites in Percut Sei Tuan, North Sumatra Province, Indonesia
**Important Value Index (IVI)**

The important value index is the mastery value of each type of vegetation in an area. Significant values are calculated with the following formula:

\[
IVI = RF + RA \text{ (for seedlings and saplings)} \tag{7}
\]

\[
IVI = RF + RA + RBA \text{ (for trees)}
\]

**Diversity Index**

Diversity \((H')\) is a parameter of vegetation that can adapt to various plant communities. In a community, there are several types of plants, therefore, the more stable the community, the higher the diversity of plant species. Species diversity can be calculated using the Shannon Index formula (Barbour et al. 1987).

\[
H' = -\sum \left( \frac{ni}{N} \ln \left( \frac{ni}{N} \right) \right) \tag{8}
\]

Information: \(H'\): Shannon Wiener’s diversity index; \(ni\): Number of individuals of a species; \(N\): Total number of individuals of a species.

Where, \(H' 0-2\): Low/little species diversity; \(H' 2-3\): Medium species diversity; \(H' > 3\): High species diversity.

The higher the value of diversity, the better the ecosystem. Conversely, it can be said that the ecosystem is extremely susceptible to pests and diseases the smaller this value is. The diversity index can also be calculated using the taxonomy index (Clarke and Warwick 2001).

\[
\Delta^+ = \frac{\sum_i \text{c}ij \cdot \text{oj}ij}{\sqrt{\text{S}(s-1)/2}} \tag{9}
\]

Where, \(\Delta^+\): Taxonomy index; \(S\): Number of species present and for double addition; \(i\) and \(j\): Range of species presence.

The taxonomic diversity index is used to identify similarities or comparisons with the Shannon-Wiener index.

**Macrozoobenthos analysis**

For determining the station, a \(1 \text{ m} \times 1 \text{ m}\) plot with a \(1 \text{ m} \times 1 \text{ m}\) horizontally arranged **paralon** pipe has been utilized. Sampling was carried out at three stations, each with three plots and each plot repeated three times. Macrozoobenthos samples can be discovered directly on the surface of the substrate, roots, and mangrove stems while looking for worms by dredging the substrate and then filtering it with a 1 mm filter eye. The water is filtered to facilitate its separation from the substrate. Furthermore, samples of macrozoobenthos were put into bottle plastics and immersed in 70% alcohol for preservation. The samples were identified, and the validity of the scientific names was further verified through the World Register of Marine Species database (http://www.marinenspecies.org), and the revised names were updated.

**Data analysis of macrozoobenthos**

**Abundance (A) and Relative Abundance (RA)**

The abundance of individual macrozoobenthos is defined as the number of species at the station in cubic units, with the formula:

\[
A = \frac{\text{Number of species (ind.)}}{\text{Sample plot area (m}^2\text{)}} \tag{9}
\]

\[
RA = \frac{ni \times 100\%}{N} \tag{10}
\]

Information: \(RA\): Relative abundance; \(Ni\): Number of individuals of type \(i\); \(N\): Total number of individuals.

**Diversity Index \((H')\), Uniformity Index \((E)\), and Dominance Index \((D)\)**

The diversity index \((H')\) provides a mathematical description of the population of organisms. The diversity of macrozoobenthos species can be calculated using the Shannon-Wiener diversity index (Odum 1993).

\[
H' = -\sum \left( \frac{ni}{N} \ln \left( \frac{ni}{N} \right) \right) \tag{11}
\]

Information: \(H'\): Shannon Wiener’s diversity index; \(ni\): Number of individuals of a species; \(N\): Total number of individuals of a species.

Where, \(H' 0-2\): Low/little species diversity; \(H' 2-3\): Medium species diversity; \(H' > 3\): High species diversity.

According to Shannon Wiener, the higher the value of diversity, the better the ecosystem. Conversely, if the diversity value is small, the ecosystem will be highly vulnerable to pests and diseases.

Uniformity is the composition of the number of individuals of each genus in the community. The uniformity index used is based on the Shannon-Wiener function to obtain the distribution of the macrozoobenthos species within the observed area (Odum 1993).

\[
E = \frac{H'}{\text{max}H'} \ln \left( \frac{1}{N} \right) \tag{12}
\]

Information: \(E\): Uniformity index; \(N\): Number of species; \(H'\): Shannon-Wiener diversity index; \(ln\): Natural logarithm; \(H'\) max: Maximum species diversity.

The value of uniformity in the population ranges from 0-1 with the following criteria (Odum 1993): \(E > 0.6\): High uniformity; \(0.4 < E < 0.6\): Medium uniformity; \(E < 0.4\): Low uniformity.

The dominance index is used to obtain information about the dominant family in a community, the calculation of which is conducted with the Simpson formula as follows (Odum 1993):

\[
D = \sum \left( \frac{P_i}{\text{Pi}} \right)^2 \tag{13}
\]

Information: \(Pi\): \(ni/N\); \(D\): Dominance index; \(ni\): Number of individuals; \(N\): Total number of individuals.

Criteria (Odum 1993), D value close to 0: No type dominates; D value close to 1: There is a dominant type.

**Measurement of physical and chemical parameters**

The physical and chemical parameters were measured in situ at the observation site, and the measurement plot was determined by sampling macrozoobenthos. The temperature is measured using a portable thermometer. Dissolved oxygen (DO) was measured using a DO meter.
(Lutron DO-5510), and pH was gauged using a pH meter (EcoTestr pH 2, Eutech). Salinity was computed using a refractometer (Atago Master 528 M), and for the chemical analysis of soil, N-Total, P-Total, and K-Total were analyzed in the PT. Socindo laboratory.

The correlation between the macrozoobenthos diversity index (H') and environmental parameters was examined using the Pearson correlation by the SPSS 21.0 software to verify the correlation value of macrozoobenthos and environmental conditions as the limiting factor (Samidurai et al. 2012). Before analyzing the correlation test as a parametric test, the data normality was tested to run normally distributed. The significance value of the variable is P < 0.05 indicates that there is a significant relationship between the two variables. If P > 0.05, there would be no significant relationship between the two variables.

RESULTS AND DISCUSSION

Vegetation analysis
At the research site, 10 true mangrove species were found, out of which seven were discovered at the seedling stage, 10 at the sapling stage, and six species at the tree stage (Table 1). Based on the research by Ningsih et al. (2011), additional species were also present, namely eight species. Avicennia marina was found to be the most common species in each growth stage, whereas Bruguiera gymnorrhiza, Ceriops tagal, Lumnitzera racemosa, and Scyphiphora hydrophyllacea were absent in the tree stage (Table 1). This study was conducted in an area bordering the sea that was overgrown by the species of A. alba, A. marina, A. officinalis, Rhizophora apiculata, and R. stylosa. This is in line with Afefe et al. (2019) study, which stated that ideal mangroves growing close to the sea are commonly covered by the Avicennia spp. and associated with Rhizophora spp in sandy or mixed soil types.

The number of individual A. marina was 548.48/ha at the seedling level, 375.76/ha for saplings, and 331.82/ha for trees. A. marina is a species that lives in wetlands and waterlogged areas. Since the area of research is close to the sea and possesses high salinity, A. marina dominates the numbers, followed by R. apiculata. A. marina has certain habitat characteristics such as high pH, temperature, thick mud, and the ability to tolerate a very low salinity range of up to 90% (Poedjirahajo et al. 2017; Balke et al. 2021). As it has breathing roots, this species can survive in places that are usually submerged in water at high tide. Overall, the number of individuals found at the study site at the tree level came up to 842.42/ha. Based on the Decree of the Minister of the Environment No. 201 of 2004 (KepMen LH No. 201 the year 2004), the standard criteria and guidelines for mangrove damage are classified as rare, or the criteria for damage is < 50%.

Table 1. Plant composition at an individual level

| Species            | Family          | Seeding | Sapling | Individual-level/Ha |
|--------------------|-----------------|---------|---------|---------------------|
| Avicennia alba     | Acanthaceae     | 183.33  | 65.15   | 75.76               |
| Avicennia marina   | Acanthaceae     | 548.48  | 375.76  | 331.82              |
| Avicennia officinalis | Acanthaceae     | 157.58  | 86.36   | 119.70              |
| Bruguiera gymnorrhiza | Rhizophoraceae   | 216.67  | 100.00  | 0                   |
| Ceriops tagal      | Rhizophoraceae  | 189.39  | 80.30   | 0                   |
| Excoecaria agallocha | Euphorbiaceae   | 0       | 57.58   | 96.97               |
| Lumnitzera racemosa | Combretaceae    | 0       | 74.24   | 0                   |
| Rhizophora apiculata | Rhizophoraceae  | 221.21  | 136.36  | 136.36              |
| Rhizophora stylosa | Rhizophoraceae  | 150.00  | 54.55   | 81.82               |
| Scyphiphora hydrophyllacea | Rubiaceae | 0       | 60.61   | 0                   |
| Total              |                 | 1666.67 | 1090.91 | 842.42              |

Table 2. Mean height and diameter at the seedling, sapling, and tree stages

| Species            | Mean height | Mean diameter |
|--------------------|-------------|---------------|
|                    | Seeding (cm) | Sapling (M)   | Tree (M) | Sapling (cm) | Tree (cm) |
| Avicennia alba     | 51.20 ± 22.51 | 3.68 ± 1.34 | 8.12 ± 4.16 | 3.50 ± 0.92 | 16.26 ± 4.39 |
| Avicennia marina   | 51.89 ± 22.95 | 3.71 ± 1.34 | 6.65 ± 3.76 | 3.58 ± 1.27 | 15.69 ± 4.94 |
| Avicennia officinalis | 51.26 ± 24.04 | 3.77 ± 1.55 | 5.68 ± 3.06 | 3.33 ± 1.23 | 16.26 ± 4.66 |
| Bruguiera gymnorrhiza | 60.27 ± 21.88 | 3.67 ± 1.50 | -         | 3.27 ± 1.11 | -          |
| Ceriops tagal      | 53.34 ± 24.38 | 4.06 ± 1.65 | -         | 3.81 ± 1.46 | -          |
| Excoecaria agallocha | -            | 3.52 ± 1.29 | 5.80 ± 2.78 | 3.24 ± 1.23 | 14.45 ± 3.44 |
| Lumnitzera racemosa | -            | 4.00 ± 1.80 | -         | 3.33 ± 1.00 | -          |
| Rhizophora apiculata | 55.01 ± 24.38 | 3.78 ± 1.42 | 7.99 ± 4.37 | 3.65 ± 1.03 | 16.31 ± 6.19 |
| Rhizophora stylosa | 52.89 ± 27.75 | 3.39 ± 0.99 | 7.80 ± 4.50 | 3.45 ± 0.92 | 17.16 ± 7.65 |
| Scyphiphora hydrophyllacea | - | 3.94 ± 1.73 | - | 3.44 ± 1.24 | - |

Note: (-) none
The species diversity index is a value that presents the diversity of the species found at the research site (Baderan et al. 2019). Specific abundance is a value proving the number of individuals of a species per unit area. The greater the density of a species, the more individual species per unit area. The highest species density was found in A. marina seedlings with a value of 1.37 ind/m²; this was due to the suitable habitat and strength of the species in adapting to its environment. Environmental conditions with sandy mud substrate can encourage the development of the A. marina species (Noor et al. 2006).

The highest species frequency was found in the growth rate of A. marina species with a value of 1.00. Species frequency describes the chance or chances of finding the species in the location area, i.e., the observation area. The R. stylosa species has a dominance value of 0.04. According to Ghufrona et al. (2015), the species dominance index is the control of a species in a location. The most important species define the control value of species in a community. The significance of a species can be used as an indicator that it is considered dominant by having a higher RD, RF, and RD value than other species (Sreelekshmi et al. 2020).

RD describes the density of the species in comparison to the whole species in an area. The species with the highest RD was A. marina at the tree level at 39.39%. This is in accordance with Parmadi et al. (2016), who declared that the high RD of A. marina is influenced by a substrate with a muddy type of sand. The highest RF value was found in the growth rate of A. marina trees with a value of 31.71% and the lowest at the sapling growth rate of A. alba with a value of 4.17%.

Dominance is the comparison between the base area and area of the plot, while RD is the overall dominance of a species. The highest RD was identified in R. stylosa (19.07%), while the lowest was in E. agallocha (8.77%). This is attributed to the fact that R. stylosa has a suitable substrate for the research location, namely sandy mud. According to Hidayatullah and Eko (2014), dominance is influenced by the same soil texture at the study site, and zoning is also included in the factors influencing the dominance of a species.

The important value index (IVI) is crucial in determining the occurrence of a species in a community. A. marina was the species with the top rank based on IVI on the seedling, sapling, and tree stages (Table 5). The IVI demonstrates a species' level of dominance, and A. marina is the most important mangrove species in the study area. Species with the greatest IVI possess better adaptability and reproductive ability compared to other species in similar habitats. The highest IVI indicates that the species can be regenerated by adapting well to the mangrove environment (Simamora et al. 2014).

Table 3. Density, frequency, dominance at the seedling, sapling, and tree stages

| Species         | Density (Individual/ha) | Frequency (%) | Dominance (m²/ha) |
|-----------------|-------------------------|--------------|------------------|
|                 | Seeding | Sapling | Tree  | Seeding | Sapling | Tree  | Seeding | Sapling | Tree  | Seeding | Sapling | Tree  |
| Avicennia alba  | 0.46    | 0.03    | 0.01  | 0.32    | 0.14    | 0.21  | 0.01    | 0.03    |
| Avicennia marina| 1.37    | 0.15    | 0.03  | 0.98    | 1.00    | 0.86  | 0.01    | 0.03    |
| Avicennia officinalis | 0.39 | 0.03    | 0.01  | 0.33    | 0.29    | 0.35  | 0.01    | 0.03    |
| Bruguiera gymnorrhiza | 0.54 | 0.04    | -     | 0.45    | 0.33    | -     | 0.01    | -       |
| Ceriops tagal  | 0.47    | 0.03    | -     | 0.38    | 0.26    | -     | 0.01    | -       |
| Excoecaria agallocha | -  | 0.02    | 0.01  | -       | 0.18    | 0.30  | 0.01    | 0.02    |
| Lumnitzera racemosa | -    | 0.03    | -     | 0.24    | -       | 0.01  | -       |
| Rhizophora apiculata | 0.55 | 0.05    | 0.01  | 0.38    | 0.39    | 0.36  | 0.01    | 0.03    |
| Rhizophora stylosa | 0.38  | 0.02    | 0.01  | 0.26    | 0.21    | 0.26  | 0.01    | 0.04    |
| Scyphiphora hydrophyllacea | -  | 0.02    | -     | 0.23    | -       | 0.01  | -       |
| **Total**      | **4.17** | **0.44** | **0.08** | **3.11** | **3.27** | **2.35** | **0.06** | **0.18** |

Note: (--) none

Table 4. Relative density (RD), relative frequency (RF), and relative basal area (RBA) at the seedling, sapling, and tree stages

| Species         | RD Seeding (%) | RD Sapling (%) | RD Tree (%) | RF Seeding (%) | RF Sapling (%) | RF Tree (%) | RBA Seeding (%) | RBA Sapling (%) | RBA Tree (%) |
|-----------------|---------------|---------------|-------------|---------------|---------------|-------------|----------------|----------------|-------------|
| Avicennia alba  | 11.00         | 5.98          | 8.99        | 10.24         | 4.17          | 9.03        | 10.18          | 17.12          |
| Avicennia marina| 32.91         | 34.47         | 39.39       | 31.71         | 30.56         | 36.77       | 10.70          | 15.95          |
| Avicennia officinalis | 9.45 | 7.92          | 14.21       | 10.73         | 8.80          | 14.84       | 9.22           | 17.12          |
| Bruguiera gymnorrhiza | 13.00 | 9.17       | -           | 14.63         | 10.19         | -           | 8.91           | -              |
| Ceriops tagal  | 11.36         | 7.37          | -           | 12.20         | 7.87          | -           | 12.11          | -              |
| Excoecaria agallocha | -           | 5.28          | 11.51       | -             | 5.56          | 12.90       | 8.77           | 13.49          |
| Lumnitzera racemosa | -             | 6.81          | -           | -             | 7.41          | -           | 9.23           | -              |
| Rhizophora apiculata | 13.27 | 12.51        | 16.19       | 12.20         | 12.04         | 15.48       | 11.10          | 17.24          |
| Rhizophora stylosa | 9.00          | 5.00          | 9.71        | 8.29          | 6.48          | 10.97       | 9.93           | 19.07          |
| Scyphiphora hydrophyllacea | -         | 5.56          | -           | -             | 6.94          | -           | 9.85           | -              |

Note: (--) none
The highest diversity index at the seedling stage was also found in *A. marina* with a value of 0.37 (Table 6). While the lowest was *R. stylosa* and *A. officinalis*, both with the value of 0.22. At the sapling stage, the highest diversity index was found in the *A. marina* species with a value of 0.37 while the lowest was in *R. stylosa* at 0.15. At the tree stage, the highest was again *A. marina* with a value of 0.37 while the lowest was *A. alba* with 0.22. Low diversity implies that the ecosystem is under pressure (Singh 2020) and in declining conditions (Macusi et al. 2021). As the mangroves live in environments with high salinity and muddy substrates, survival only occurs for those species with high adaptation abilities after a strict selection (Noor et al. 2006).

The level of mangrove diversity in Percut Sei Tuan is classified into low criteria in seedlings and trees, while saplings are classified into the moderate criteria (Barbour et al. 1987). The low diversity of mangrove species in the study area might be attributed to environmental conditions that are only able to support the growth of certain species or because the mangrove species utilize their greater regeneration power.

The highest taxonomic diversity index for the seedling stages was confirmed to be in the *A. marina* species with a value of 0.16 and the lowest was in *R. stylosa* (Table 7). At the sapling stage, *A. marina* still owned the highest value at 0.17 and the lowest with a value of 0.02 was found in *R. stylosa*. At the tree stage, once again, *A. marina* dominated with a value of 0.20, while *R. apiculata* and *A. alba* had the lowest value of 0.04. The diversity index in a community greatly influences the number of individuals and species in the community. Eddy et al. (2021) claimed that species diversity in a community would be high if the community consisted of several species and no species was dominant. Conversely, if a community has a low diversity value, then it will consist of species and additionally, there will be a dominant species.

Identification of macrozoobenthos

Four classes of the macrozoobenthos were found in the Percut Sei Tuan mangrove forest in North Sumatra: Mollusca, Gastropoda, Arthropoda, and Polychaete. This finding is rather similar with previous research conducted by Basyuni et al. (2018), who found three classes of macrozoobenthos, namely Gastropoda, Bivalves, and Malacostraca. The species most commonly found are from the class Gastropoda with 9.44 ind/m² (Table 8). It has a high tolerance to environmental changes, as well as high adaptability and a wide distribution range. Polychaeta, consisting of solely one species, might be caused by its slow response-ability. This is in accordance with Kasumyan (2019), who stated that in certain soft-bodied animals, such as Polychaeta worms, osmotic regulation mechanisms were developed, but the response was relatively slow.

**Table 5. Important value index at seedling, sapling, and tree stages**

| Species              | Important value index (IVI) |
|----------------------|-----------------------------|
|                      | Seedling (%) | Sapling (%) | Tree (%) |
| Avicennia alba       | 21.24         | 20.32       | 35.15    |
| Avicennia marina     | 64.62         | 75.73       | 92.12    |
| Avicennia officinalis| 20.19         | 25.94       | 46.17    |
| Bruguiera gymnorrhiza| 27.63         | 28.27       | -        |
| Ceriops tagal        | 23.56         | 27.35       | -        |
| Excoecaria agallocha | -             | 19.61       | 37.90    |
| Lumnitzera racemosa  | -             | 23.45       | -        |
| Rhizophora apiculata | 25.47         | 35.65       | 48.91    |
| Rhizophora stylosa   | 17.29         | 21.42       | 39.75    |
| Scyphiphora hydrophyllacea | -           | 22.35       | -        |

*Note: (-) not detected*

**Table 6. Shannon Wiener diversity index at seedling, sapling, and tree levels**

| Species              | Diversity index (H') |
|----------------------|----------------------|
|                      | Seedling | Sapling | Tree  |
| Avicennia alba       | 0.24     | 0.17    | 0.22  |
| Avicennia marina     | 0.37     | 0.37    | 0.37  |
| Avicennia officinalis| 0.22     | 0.20    | 0.28  |
| Bruguiera gymnorrhiza| 0.27     | 0.22    | -     |
| Ceriops tagal        | 0.25     | 0.19    | -     |
| Excoecaria agallocha | -        | 0.16    | 0.25  |
| Lumnitzera racemosa  | -        | 0.18    | -     |
| Rhizophora apiculata | 0.27     | 0.26    | 0.29  |
| Rhizophora stylosa   | 0.22     | 0.15    | 0.23  |
| Scyphiphora hydrophyllacea | -      | 0.16    | -     |
| **Total**            | **1.83** | **2.06**| **1.63** |

*Note: (-) none*

**Table 7. Taxonomic index at seedling, sapling, and tree levels**

| Species              | Taxonomic diversity index |
|----------------------|---------------------------|
|                      | Seedling | Sapling | Tree  |
| Avicennia alba       | 0.05     | 0.03    | 0.04  |
| Avicennia marina     | 0.16     | 0.17    | 0.20  |
| Avicennia officinalis| 0.05     | 0.04    | 0.07  |
| Bruguiera gymnorrhiza| 0.06     | 0.05    | -     |
| Ceriops tagal        | 0.06     | 0.04    | -     |
| Excoecaria agallocha | -        | 0.03    | 0.06  |
| Lumnitzera racemosa  | -        | 0.03    | -     |
| Rhizophora apiculata | 0.07     | 0.06    | 0.08  |
| Rhizophora stylosa   | 0.04     | 0.02    | 0.05  |
| Scyphiphora hydrophyllacea | -      | 0.03    | -     |
| **Total**            | **0.50** | **0.49**| **0.49** |

*Note: (-) none*
Table 8. Identification of macrozoobenthos

| Class   | Family   | Species              | Station | Individual-level |
|---------|----------|----------------------|---------|-----------------|
| Mollusc | Ellobiidae | Cassidula angulifera | 1       | +               |
| Mollusc | Ellobiidae | Cassidula nucleus    | 2       | +               |
| Gastropod | Potamididae | Cerithidea obtusa | 3       | +               |
| Gastropod | Ellobiidae | Ellobium aurisjudae |         | +               |
| Arthropod | Sesarmidae | Episesarma versicolor |    | +               |
| Polychaete | Nereididae | Nereis sp            |         | +               |
| Arthropod | Ocyopodidae | Uca tetragonon |         | -               |

Note: (-) none

Table 9. Individual-level and ecological index of macrozoobenthos at each research station

| Station/transect | Individual-level ind/m² | Diversity index (H) | Ecological index value | Dominance index (D) |
|------------------|--------------------------|---------------------|-----------------------|---------------------|
| 1                | 9.33                     | 1.81                | 0.93                  | 0.19                |
| 2                | 7.44                     | 1.27                | 0.71                  | 0.34                |
| 3                | 8.78                     | 1.54                | 0.86                  | 0.25                |

The diversity index value of macrozoobenthos ranges from 1.54 to 1.81, which is classified into the low to a medium category, while the uniformity index value ranges from 0.71 to 0.93 and is included in the high category (Table 9). The dominance index value ranges from 0.19 to 0.34 and belongs in the non-dominant category. The indexes of diversity, uniformity, and dominance are widely used in evaluating the aquatic environment based on its biological conditions. The unbalanced environmental conditions will affect organisms living in the waters (Odum 1993). The diversity index value has the highest value at the station 1 at 1.81 and is grouped into the moderate category. Theuerkauff et al. (2018) reported that species diversity is the variation in an ecosystem; when it has a high diversity index, the ecosystem tends to be balanced and vice versa.

The uniformity index displayed the highest value at the station 1 at 0.93 and fell under the high category. According to Gravina et al. (2020), the uniformity index determines if there exists a dominant pattern of one or many species groups in the observed environment. If the uniformity index value is equal to one, the distribution of individuals among species is quite even. If it is close to 0, there are certain groups of species that are relatively more abundant than other species. The dominance index in station II also exhibits a high value (0.34), so it can also be classified as a dominant species. This is in accordance with Odum (1993), who expressed that the dominance index value varies from 0 to 1; if the dominance index value is close to 0, no individual species dominates; conversely, if the dominance index value is close to 1, there is a dominant individual. A high dominance index value indicates a high concentration of dominance, and a low value, on the other hand, indicates a low concentration.

Relationship between macrozoobenthos diversity and environmental parameters

The Pearson correlation with SPSS 21.0 software has been employed in examining the relationship between the value of the Macrozoobenthos Diversity Index (H) and environmental conditions (Table 10). The presence of macrozoobenthos in mangrove waters is strongly influenced by various environmental factors: temperature, pH, salinity, and DO (Harahap et al. 2018).

Table 10. Relationship of macrozoobenthos diversity and environmental parameters using Pearson correlation

| Parameter         | P value | Pearson correlation | Correlation type |
|-------------------|---------|---------------------|------------------|
| Salinity (%)      | 0.31    | 0.38                | Low              |
| pH                | 0.37    | 0.34                | Low              |
| Dissolved oxygen (DO) (mg/l) | 0.045 | 0.68                | High             |
| Temperature (˚C) | 0.39    | 0.33                | Low              |

Table 11. Physical and chemical habitat characteristics of mangroves at each station

| Station | Salinity (%) | pH | Dissolved oxygen (mg/l) | Temperature (˚C) |
|---------|--------------|----|-------------------------|-----------------|
| 1       | 23.67 ± 1.53 | 6.17 ± 0.29 | 7.37 ± 2.06 | 32.70 ± 2.11 |
| 2       | 21.67 ± 2.88 | 6.00 ± 0.26 | 6.13 ± 2.06 | 31.03 ± 2.08 |
| 3       | 26.33 ± 8.50 | 6.10 ± 0.10 | 8.17 ± 2.50 | 30.23 ± 2.58 |
Table 12. Characteristic of soil chemical

| Station | N-Total (%) | P-Total (%) | K-Total (%) |
|---------|-------------|-------------|-------------|
| 1       | 0.32 ± 0.07 | 0.08 ± 0.00 | 0.52 ± 0.03 |
| 2       | 0.36 ± 0.06 | 0.08 ± 0.01 | 0.43 ± 0.07 |
| 3       | 0.37 ± 0.04 | 0.09 ± 0.01 | 0.44 ± 0.02 |

The availability of the DO parameter with a value of 0.68 is strongly correlated with macrozoobenthos diversity in mangrove waters, so it can be assumed that DO affects the presence of macrozoobenthos in mangrove waters. The highest DO value was found at station III, which was 8.17 mg/l. The higher the DO concentration in the waters, the higher its quality and vice versa, which might impact the existence of macrozoobenthos (Sahidin et al. 2018). Moreover, Harahap et al. (2018) presented that DO is positively correlated with macrozoobenthos and may indeed be employed as a bioindicator as well as a determinant of water quality.

The relationship between the number of macrozoobenthos individuals in each observation plot and environmental parameters was positive because its value was not equal to 0. Of the four environmental parameters observed, only the DO value is significant because significant < 0.05 with macrozoobenthos diversity, the rest have a weak type of correlation, which is depicted in Table 10. The temperature in this study owns a value of 0.33 and has a low correlation, but this value will still be tolerated and accepted by macrozoobenthos in accordance with the quality standard of the Decree of the Minister of the Environment No. 51 (2004). The temperature for marine biota is 28-32°C, whereas the temperature at the research site ranges from 29-33°C. Macrozoobenthos can grow optimally with pH values ranging from 5-9; overall from the parameters observed, it is known that these waters are still considered feasible to support macrozoobenthos survival (Basyuni et al. 2018; Susetya et al. 2018).

Station III has the highest N-Total (0.37%) value due to its abundance of vegetation and marine life. Consequently, N has a major impact on plants and the availability of other nutrients. Soil nitrogen, a critical nutrient for plants, is required for plant vegetative development as well as protein synthesis (Nugroho et al. 2013). Station III has the highest Phosphorus (P-Total) value of 0.09%, which is caused by the soil’s acidic pH and the presence of organic materials from decomposition. Activity on lands, such as natural environmental impacts or human activities, can alter the availability of a soil element, one of which is phosphorus (Holliday and Gartner 2007). Potassium (K-Total) has the greatest value of 0.52% at a station I as it is the macronutrient, after nitrogen, that is most widely absorbed by plants; nevertheless, in the results gathered at the research location at a station I, it is higher so that the nitrogen nutrients can be absorbed into the plants. Potassium is absorbed by plants in amounts equivalent to or greater than nitrogen. If the potassium in the soil is adequate for plant growth, the plant will grow well (Etesami et al. 2017).

Discussion

The vegetation analysis in the Percut Sei Tuan mangrove forest, North Sumatra, resulted in the discovery of 10 species, namely Avicennia alba, A. marina, A. officinalis, Bruguiera gymnorrhiza, Ceriops tagal, Excoecaria agallocha, Lumnitiera racemosa, R. apiculata, R. stylosa, and Scyphiphora hydrophila. The species A. marina was the most common one found at each growth stage, while A. alba was the lowest at the tree stage and R. stylosa in the sapling stage. At the study location study, an area bordering the sea, there are species of A. alba, A. marina, A. officinalis, R. apiculata, and R. stylosa. This finding followed Afef et al. (2019), who reported that the characteristics of ideal mangroves that grow in areas close to the sea are often overgrown by Avicennia spp. and associated with Rhizophora spp. in sandy or mixed soil types. Recently, this finding was supported by Balke et al. (2021) as this location faces the open coast directly, i.e., faces seaward, colonizing A. marina.

The macrozoobenthos found in the Percut Sei Tuan mangrove forest, North Sumatra, consisted of two types of mollusks, Cassidula angulifera and Cassidula nucleus, two species of Gastropoda, Cerithidea obtusa and Elobium aurisjudae, two species of Arthropoda, Evisarma versicolor and Eca tetragonon, and one species of Polychaeta, Neris sp. The species most commonly found are from the class Gastropoda, which has a high tolerance to environmental changes, as well as high adaptability and a wide distribution range. This is supported by Silaen et al. (2013) who confirmed that the density and distribution of Gastropoda depend on their environment and habitat, the presence of food, ecological pressures, and environmental changes in mangrove vegetation.

The number of individuals in the Percut Sei Tuan mangrove forest is 842.42/ha. Based on the Decree of the Minister of the Environment No. 201 of 2004 (KepMen LH No. 201 of 2004), the standard criteria and guidelines for mangrove damage are classified as rare or the criteria for damage are < 50%. While the diversity index (H') of macrozoobenthos ranged from 1.54 to 1.81 ind/m², which was included in the low to a medium category, mangroves with a high density provide good shelter and support the availability of adequate nutrient intake from litter (Bayan et al. 2016). Fallen mangrove leaves will be decomposed by microbial communities that engage in further functions to produce organic matter and nutrients in the soil, land, and ecosystem (Karniati et al. 2021). This relationship maintains the ecosystem’s balance in the substrate and serves as a food supply for macrozoobenthos. As per the research conducted by Nugroho et al. (2013), the higher the mangrove density, the more litter produced, which enhances the organic matter content in the substrate. The survival of marine species, particularly macrozoobenthos, will be damaged if the mangrove ecosystem is disturbed. Our findings pointed out that it is necessary to carry out rehabilitation efforts that involve the community and mangrove area management sector in a sustainable manner in the Percut Sei Tuan mangrove forest area. The impact of the destruction of the mangrove ecosystem will be very large, both from an ecological and economic perspective.
(Su et al. 2021). The marine food chain will be destroyed, leading to a decline in the fisheries sector (Macusi et al. 2021). The coastal community’s protection from sea-level rise, storm surges, and tsunamis will be affected (Maraio et al. 2015). Human communities living in or near mangroves will lose access to essential food, fiber, timber, chemicals, pharmaceuticals, animal feed, and fertilizers (Van Oudenhoven et al. 2015; Sari et al. 2018; Rupidara et al. 2020).

In conclusion, the current study found 10 true mangrove species where *Avicennia marina* was the most abundant species at each seedling, sapling, and tree level, indicating its highly relative species diversity. The physical and chemical factors of the observed parameters suggested that these mangrove waters are still considered suitable for supporting the lives of macrozoobenthos with dissolved oxygen, which was strongly correlated to macrozoobenthos diversity. Further research is required to continue exploring the diversity of macrozoobenthos on the restored mangroves.

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