The Toxic Effect of Conventional Treated Mine Water on Aquatic Organisms

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The present paper aims to evaluate the toxic effects generated by conventional treated mine water as well as freshwater samples (Macris river) on primary producers (green algae, Selenastrum capricornatum) and primary consumers (planktonic crustaceans, Daphnia magna). Those organisms could be very reliable biological models to assess the toxic effect of mine waters on the environment. The green algae growth inhibition test (performed according to OECD201) and the acute immobilization test of freshwater crustaceans (Daphnia magna) (performed according to OECD 202) showed no growth inhibition or Daphnia magna immobilization / mortality during 72h and 48h incubation in presence of surface water samples from Macris river. The aqueous system with 1.89 g/L sulphate ion (SO\text{4}^{2-}) concentration, containing treated mine water mixed with freshwater sample (ratio 2.4:1), generated 100% toxic effect on crustaceans (Daphnia magna) after 48h incubation.

Keywords: conventional treated mine water, ecotoxicity, Selenastrum capricornatum, Daphnia magna

The bioassays (using biological models such as bacteria, fungi, yeasts, microalge, plants, invertebrates and vertebrates) have been designed in order to have an extensive repertoire of responses (growth, mortality, reproduction) when they are exposed to specific stressors (chemical compounds and / or contaminated environmental samples) [1, 2]. The response (test endpoint) is based on the organisms sensitivity and adaptability [3].

The REACH Regulation on the Registration, Evaluation, Authorization and Restriction of Chemicals requires the reduction of conventional toxicity tests (applied to vertebrate organisms) and their replacement with alternative toxicity tests using microbiotests. In addition, comparative acute toxicity studies using conventional biological models (e.g. fish) and alternative toxicity tests (on algae and crustaceans) showed similar sensitivity, which leads to the hypothesis that microbiotests are cost-effective and a valid alternative for assessing the toxic effects generated by chemical compounds / pollutants [4-8].

However, a disadvantage of using microbiotests is that they do not identify the nature of the chemical / pollutant in water, but indicate the presence of a toxic substance [9]. Generally, the evaluation criterion for toxicity tests is the percentage of inhibition / stimulation (in the case of tests using algae, bacteria, protozoa) or immobilization / mortality (in tests using crustaceans, rotifers).

The results of these toxicity tests are expressed by the actual concentrations at which mortality / immobilization can be recorded for 50% of the tested organisms (EC50). Another important criterion is the sensitivity. The bioassay should provide a quantifiable result for low concentrations of pollutant that is within or below the concentration that may affect the environment and human health [10, 11].

Mine water conventional treatment is done by pH adjustment to 8.5-10.5 with lime or calcite which efficiently removes the heavy metals by precipitation as hydroxylated species. The process can also lower the sulphate to a concentration of minimum 1500 mg/L, corresponding to the gypsum solubility. Unfortunately, this treatment process generates a high calcium sulphate residual concentration (1500-4000 mg SO\text{4}^{2-}/L), because gypsum precipitation is limited by thermodynamic and kinetic factors. Gypsum crystallization can be improved by seeding it with recycled gypsum in a separate treatment stage, but this is not common for acid mine water treatment [12]. While heavy metal species precipitation can be properly done, the calcium sulphate remaining dissolved in effluent can have a negative impact on surface water biocenosis.

In this study, the assessment of the toxic effects generated by a simulated mixture of conventionally treated mine water and surface water from Macris river as well as Macris river water (upstream of mine water confluence) was performed on two biological models belonging to two trophic levels, primary producers (green algae, Selenastrum capricornatum) and primary consumers (planktonic crustaceans, Daphnia magna), given that nowadays the mine water is discharged untreated into the Macris freshwater system.

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Experimental part

The mine water from P2 source Coranda, Certej was treated in laboratory by a conventional process: precipitation with calcium hydroxide, pH = 9.5 and a reaction time of 30 min. The solids were separated by settling, resulting a conventional mine water treated as effluent $E1$. A seeded gypsum crystallization was further applied, in order to force the system to advance toward equilibrium, using 15 g/L gypsum for seeding and a 30 minutes contact time, obtaining a second mine water effluent $E2$.

A mixture (Mix1) was made using conventional treated mine water $E1$ and surface water from Macris river (mixing ratio of 2:4:1) to simulate the conditions that would occur at aquatic system level by discharging the effluent ($E1$) to the Macris river.

The mixture ratio of 2:4:1 was established based on the most unfavourable real life values recorded by Cepromin S.A, the most favorable ratio being 1.9:1. The toxic effects generated by the conventional treated mine water and surface water (Mix 1) were assessed based on the sensitivity of aquatic organisms.

Biological models used in the toxicity tests.

*AlgalToxkit™ bioassay.* Green algae inhibition assay (AlgalToxkit™, batch no. SC200218, Microbiotests Inc., Belgium) was performed following OECD 201 (similar to C3 method published in Annex C, Regulation 440/2008) and ISO 8692:2012 requirements [13]. The microalgae with an exponential growth curve were incubated for 72h in presence of conventional treated mine water and surface water in 10 cm test cuvettes. The water samples did not require the addition of nutrients. The test solutions were prepared prior to the start of the test by diluting the water sample with culture medium up to an initial concentration of $1\times10^4$ algae cells / ml. Serial dilutions (three replicates) were performed as follows: C1-100%, C2-50%, C3-25%, C4-12.5%, C5-6.25% in 200 ml volumetric flask. The test endpoint was the inhibition and / or stimulation of algal growth after 72h. After homogenization of the obtained mixtures, 25 ml of each test solution (C1-C5) as well as control sample (C0) was transferred into 3 test cuvettes for each concentration. Samples were incubated at 22-25°C and constant illumination (Aqualytic, Microbiotests Inc., Belgium). Growth or inhibition of algal cells was quantified by measuring the optical density at 670 nm after 24h, 48h and 72h incubation (VWR UV-VIS Spectrometer, USA).

*DaphToxkit™ bioassay.* The acute immobilization test of freshwater crustaceans (*Daphnia magna*, DaphToxkit™, batch no. DM231117 (Microbiotests Inc., Belgium) was performed following OECD 202 guideline and Method C.2 (Annex C, Reg. CE 440/2008) and ISO 6341: 2013 [14]. *Daphnia magna* ephippia, prior to the start of the toxicity test, were incubated for three days at 20-22°C and a light intensity of 6000 lux (Aqualytic, Microbiotests Inc., Belgium) for neonatal hatching. The neonates before testing were subjected to a 2-hour pre-feeding procedure with microalgae. Similar to the green algae inhibition test, daphnia neonates were exposed for 48 hours to water samples (conventional treated mine water and surface water) at different concentrations (C1: 100%, C2: 50%, C3: 25%; C4: 12.5%; C5: 6.25%). The tests were performed in plates using 20 neonates of *Daphnia magna* for each test solution (C1-C5) and control sample (C0) (only culture medium with neonates). The endpoint of the test was immobilization and / or mortality of crustaceans after both 24h and 48h incubation time, respectively. Four replicates of each concentration were tested (5 concentration x 4 replicate).

MBT Regetox 2.0 software (Microbiotests Inc., Belgium) was used for experimental data interpretation.

Results and discussions

The conventional effluent ($E1$) had low metal concentrations but it was supersaturated with $\text{CaSO}_4$ ($\text{SO}_4^{2-}$ approx. 3000 mg/L), because of limited calcium sulphate precipitation and crystallization as gypsum. The secondary effluent $E2$, from the seeded gypsum crystallization had an improved quality, with a final sulphate concentration c($\text{SO}_4^{2-}$) = 1739 mg/L. Analytical results showed major differences between $E1$ and $E2$ in terms of $\text{SO}_4^{2-}$ and Ca$^{2+}$. A decrease of both $\text{SO}_4^{2-}$ as well as Ca$^{2+}$ was observed, 1.29 g/L in case of $\text{SO}_4^{2-}$ and 0.52 g/L for Ca$^{2+}$, respectively (Table 1).

| Table 1 | THE RESULTS OF PHYSICAL-CHEMICAL PARAMETERS IN MINE WATER, TREATED MINE WATER AND SURFACE WATER FROM P2 SOURCE, CORANDA, CERTEJ |
|-----------------|------------------|-----------------|-----------------|-----------------|-----------------|
| Indicator       | U.M              | Mine water      | Effluent $E1$   | Effluent $E2$   | Macrisului River | Mix 1           |
| pH              | 2.69             | 9.5             | 8.57            | 7.65            | 8.16            |
| Electric conductivity | mS/cm     | 3.90             | 3.58            | 2.17            | 0.201           | 2.84            |
| Ca(OH)$_2$ consumption | g/L       | -               | 3.1             | 0               | -               | 0               |
| Gypsum dose     | g/L              | -               | -               | 15              | -               | -               |
| Reaction time   | min              | -               | 30              | 30              | -               | -               |
| Reaction temperature | °C       | 23              | 24.7            | -               | -               | -               |
| $\text{SO}_4^{2-}$ | mg/L          | 3532            | 2998            | 1739            | 26              | 1890            |
| Ca$^{2+}$       | mg/L            | 313             | 1281            | 763             | 3.7             | 715             |
| Mg$^{2+}$       | mg/L            | 311             | 43              | 57.3            | 5.7             | 41              |
| Total dissolved solids | mg/L      | 5360            | 4382            | 2615            | 118             | 2746            |

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Fe | mg/L | 228 | 0.098 | <0.002 | 0.115 | -
Al | mg/L | 167 | 1.84  | 1.59  | -     | -
As | mg/L | 0.041| <0.005| <0.005| 0.0055| -
Cu | mg/L | 4.59 | <0.001| <0.001| 0.0045| -
Zn | mg/L | 185 | 0.005 | 0.006 | 0.0067| -
Mn | mg/L | 131 | 0.255 | 0.197 | 0.0012| -
Cd | mg/L | 1.07 | 0.002 | 0.002 | 0.0005| -
Pb | mg/L | 0.28 | <0.006| <0.006| <0.005| -
Co | mg/L | 1.06 | 0.001 | 0.002 | <0.0008| -
Cr | mg/L | 0.031| <0.001| 0.002 | 0.0136| -

Toxic effects of the treated mine water $E1$ on freshwater organisms were possible due to heavy metal species or to salinity, mainly calcium sulphate in solution. However, heavy metal concentration values were low after treatment, so a special focus was kept on sulphate ion as a key parameter.

The water mixture $Mix1$, as well as surface water samples from Macris river were tested on different biotic components of aquatic systems biocenosis (green algae and planctonic crustaceans). Tests were performed to assess the degree of inhibition of green algae growth after a 72h incubation and planktonic crustacean immobilization / mortality after exposure for 48 hours in the presence of water samples.

Assessment of green algae ($Selenastrum capricornutum$) growth inhibition.

The toxicity assessment of the mixture containing conventional treated mine water ($Mix 1$) revealed no inhibition effect of $Selenastrum capricornutum$ growth (Fig. 1).

On the contrary, a stimulation of green algae growth was observed. The optical density measured at 670 nm showed that $Selenastrum capricornutum$ growth steadily increased during the incubation time.

The highest algal growth was recorded after 72h at C3 (25%) solution followed by C4 (12.5%) corresponding to 0.47 g / L and 0.24 g / L $SO_4^{2-}$, respectively.

In addition, no inhibitory effect on the species tested ($Selenastrum capricornutum$) was recorded in case of the freshwater Macris river (data not shown).

Assessment of planktonic crustaceans ($Daphnia magna$) immobilization / mortality.

The toxicity assessment of conventional treated mine water samples (mixture of treated by mine water and surface water Macris river, $Mix 1$) on aquatic crustaceans after 24h showed no toxic effect (Table 2 and Figure 2a). Nevertheless, 100% mortality of crustaceans was detected for the tested water sample (undiluted), C1 (100%).

### Table 2

| Simul 1 | $R_1$ | $R_2$ | $R_3$ | $R_4$ |
|---------|-------|-------|-------|-------|
| Conc. (%) | Effect% after 24h | | | |
| 100 | 5 | 10 | 0 | 15 |
| 50 | 0 | 0 | 15 | 5 |
| 25 | 0 | 0 | 0 | 0 |
| 12.5 | 0 | 0 | 0 | 0 |
| 6.25 | 0 | 0 | 0 | 0 |

Note: R - replicate
The percentage of immobilization / mortality of *Daphnia magna* decreased considerably from 10-20% for the tested C2 solution (50%), to 5% for the tested solution C3 (25%) and 0% for the following tested solutions (C2 = 12.5% and C1= 6.25%) (Table 2).

Considering all the test solutions, the effective concentration (EC50) at which the conventional treated mine water (*Mix 1*) samples inhibited 50% of *Daphnia magna* species was 63.5% (Fig. 2b).

Thus, considering that the initial SO$_4^{2-}$ concentration of 1890 mg/L of *Mix 1* and EC50 calculated after 48h incubation in presence of *Daphnia magna*, we conclude that a concentration of SO$_4^{2-}$ of 1200 mg/L can induce immobilization and/or mortality of 50% of *Daphnia magna* after 48h exposure.

Moreover, the effects of freshwater samples from Macris river on planktonic crustaceans (*Daphnia magna*) were not significant (0% immobilization and/or mortality). Similar results were confirmed by van Dam et al. (2014) [15] on *Moinodaphnia macleayi* (crustaceans) and *Chlorella sp.* (green algae) regarding the sensitivity of crustaceans species to that of green algae after exposure to wastewaters containing sulphates. In addition, van Dam et al. (2014) [15] showed that the main specific ions do not clearly cause the toxicity of mine waters, but the effects are due to electrical conductivity. Various toxicity data identified in literature (Table 3) showed comparable data with our present study.

### Table 3

| Toxicity data identified in the literature | EC50 values detected in the study |
|-------------------------------------------|----------------------------------|
| **CaSO$_4$**                             |                                  |
| **ALGAE**                                |                                  |
| *Selenastrum capricornutum*              |                                  |
| EC50=72h / inhibitie sau stimulare a cresterii algelor verzi | EC50>79 mg/l (MSDS Reach Regulation) |
| Nitzchia linearis                        |                                  |
| LC50 (5 days) = 3.200.000 µg/l           |                                  |
| PAN Pesticide Database –Chemicals        |                                  |
| *Chlorella vulgaris*                     |                                  |
| NR (30 days) = 1.872.000 µg/l            |                                  |
| 1.497.000µg/l                            |                                  |
| PAN Pesticide Database –Chemicals        |                                  |
| *Navicula seminulum*                     |                                  |
| EC50 (96h) = 3.200.000 µg/l              |                                  |
| PAN Pesticide Database –Chemicals        |                                  |
| **CRUSTACEA**                            |                                  |
| *Daphnia magna*                          |                                  |
| EC50=48h /immobilization / mortality     |                                  |
| LC50(48h) = 1.470.270µg/l                |                                  |
| 859.960 µg/l                             |                                  |
| 1.470.240 µg/l                           |                                  |
| PAN Pesticide Database –Chemicals        |                                  |
| Ceriodaphnia dubia                       |                                  |
| LC50(24h, 48h) = 1.940.000 µg/l          |                                  |
| PAN Pesticide Database –Chemicals        |                                  |

EC50-48h = 60-65%  
EC50-48h = 1.2 g/L SO$_4^{2-}$ concentration

[15] van Dam et al. (2014)
Conclusions

The evaluation of toxic effects of water samples (simulated conventional treated mine water and freshwater from Macris river) on aquatic organisms (*Selenastrum capricornutum* and *Daphnia magna*) revealed that:

- in terms of **physical-chemical** analysis, the conventional treatment was effective in removing the heavy metals, but with limited efficiency for sulphate anions;
- in terms of **toxicity assessment**, the sulphates and the remaining heavy metals could have a toxic effect on freshwater trophic chain;
- the surface water of Macris river did not induced either growth inhibition of green algae, nor immobilization / mortality of planktonic crustaceans;
- the conventional treated mine water (mixture of P2 source, Coranda, Certej and Macris river, Mix1 with SO$_4^{2-}$ concentration of 1.89 g/L) stimulated the green algae growth after 72h incubation time, especially for 25% and 12.5% test solutions;
- the mixture between treated mine water and Macris river, Mix1 with 1.89 g/L SO$_4^{2-}$ concentration generated a toxic effect (immobilization / mortality) of 100% on crustaceans (*Daphnia magna*) after 48h incubation. The EC50 value at 48h was 63.5%, corresponding to 1.2 g/L SO$_4^{2-}$ concentration;
- the water mixture with a SO$_4^{2-}$ content of up to 1.2 g/L does not cause immobilization / mortality of *Daphnia magna* planktonic crustaceans;
- the conventional treated mine water discharge with 1.89 g/L SO$_4^{2-}$ concentration in an aquatic system (either lake or river) may influence the eutrophication process in terms of augmentation.

Overall, an additional seeded gypsum crystallization treatment phase of mine water treatment (effluent type E2), had a significant reduction effect of SO$_4^{2-}$ (42%) as well as Ca$^{2+}$ (59%). The SO$_4^{2-}$ output to the river could be reduced by 42%, which means a reduction of the sulphate concentration in the river after mixing of about 34% (e.g. a maximum of 1.24 g SO$_4^{2-}$/L for the unfavorable mixing ratio and 1.15 g of SO$_4^{2-}$/L for the more favorable blend ratio). Thus, the simple use of the gypsum seeded crystallization phase, desaturation of effluent in addition to conventional treatment, is likely to bring a favorable change in terms of the water body ecology.

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