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Exposure of Metal Oxide Nanoparticles on the Bioluminescence Process of P<sub>m</sub>- and P<sub>n</sub>-lux Recombinant P. putida mt-2 Strains

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Abstract: Comparison of the effects of metal oxide nanoparticles (NPs; CuO, NiO, ZnO, TiO<sub>2</sub>, and Al<sub>2</sub>O<sub>3</sub>) on different bioluminescence processes was evaluated using two recombinant (P<sub>m</sub>-lux and P<sub>n</sub>-lux) strains of Pseudomonas putida mt-2 with same inducer exposure. Different sensitivities and responses were observed according to the type of NPs and recombinant strains. EC<sub>50</sub> values were determined. The negative effects on the bioluminescence activity of the P<sub>n</sub>-lux strain was greater than for the P<sub>m</sub>-lux strains for all NPs tested. EC<sub>50</sub> values for the P<sub>m</sub>-lux strain were 1.7-to 6.2-fold lower (corresponding to high inhibition) than for P<sub>n</sub>-lux. ZnO NP caused the greatest inhibition among the tested NPs in both strains, showing approximately 11 times less EC<sub>50</sub>s of CuO, which appeared as the least inhibited. Although NPs showed different sensitivities depending on the bioluminescence process, similar orders of EC<sub>50</sub>s for both strains were observed as follows: ZnO > NiO, Al<sub>2</sub>O<sub>3</sub> > TiO<sub>2</sub> > CuO. More detailed in-depth systematic approaches, including in the field of molecular mechanisms, is needed to evaluate the accurate effect mechanisms involved in both bioluminescence metabolic processes.

Keywords: bioluminescence; inducer; nanoparticles; P<sub>m</sub>- and P<sub>n</sub>-lux; recombinant

1. Introduction

Petroleum-based hydrocarbons are an important source of energy and are also used as raw materials for the production of various organic compounds, but they are becoming some of the most important and common pollutants in ecosystems [1–3]. Although a large number of remediation technologies are generally applied to clean up such contaminated sites, biomonitoring may be an effective remediation strategy in some cases, especially for groundwater and subsurface contamination where contaminant plumes are retained within sites [4]. This requires a variety of biomonitoring technologies including bioassays and biosensors, which provide information about the impact of contaminants on organisms and ecosystems [5,6]. Various microorganisms and higher life (i.e., fishes, plants, and invertebrates) are used for biomonitoring, while microbial processes are especially attractive for contaminated sites [7,8]. Among various microbial processes, methods using the activity of whole cell, specific enzymes, and intact or recombinant gene expression have been adopted for assessing the environmental pollutants by regulatory and monitoring agencies because they are simple, rapid, and low cost. These also can be used for preliminary assessment and proper complementary approaches for interpretation of the chemical and physical analysis [4,9]. Among the applicable microbial activities, recombinant genes and their regulation have resulted in interest in tools for the development of environmental monitoring. The lux genes found in Vibrio fischeri (luxCDABE) (recently named as Aliivibrio fischeri) is one of the most applicable reporter genes [4]. Recombinant
strains containing lux genes can be used for real-time, non-destructive assays based on bioluminescence (visible light) production. Bacteria containing recombinant lux-genes and the TOL plasmid, that encodes enzymes converting toluene analogs (notably gasoline components) into non-toxic intermediates, have been shown to be effective bioindicators of contaminants [10,11]. Bioluminescence is related to bacterial metabolic activity, and could be a sensitive indicator of bacterial metabolic status [4]. When these lux genes combined with other genes that are activated in the presence of specific contaminants, the lux gene sequence of V. fischeri acts as an effective bioreporter for the applications of environmental biomonitoring. One example of bioluminescence induction is P. putida mt-2 KG1206 and RB 1401, which contain the $P_m$-lux and $P_u$-lux recombinant plasmid with TOL plasmid, respectively. Two regulatory genes, XylS and XylR control the catabolic activity of the TOL plasmid genes, and these two genes act as transcriptional activators on two operons of the upper and lower pathways [10,11].

The two strains used for this investigation can produce bioluminescence in the presence of toluene analogs and their intermediates, and therefore, can be used for preliminary biomonitoring approaches of the gasoline-contaminated sites. Valuable information about the bioavailability of contaminants in the environment can be generated by these bioluminescence activities. These results can serve as useful alternatives compared to expensive analytical methods for on-site analysis or in situ monitoring.

Biomonitoring studies have generally focused on individual contaminants under laboratory conditions. However, the contaminated site is generally exposed to a mixture of pollutants rather than one. Therefore, analyzing mixtures will portray the realistic conditions of biomonitoring sites. The commercial products of NPs are increasingly found in the environment due to the fast growing nanotechnology field, a consequence of their global economic importance. They are one of the possible co-pollutants with hydrocarbons [12–14]. NPs technologies have progressed due to their unique and distinctive characteristics (high specific surface area, small size distribution, high dispensability, etc.) and are entering many industrial areas such as electronics, energy, medicine, plastics, and aerospace [15]. The promise of any new technology, however, must be weighed against the possibility of generating exceptional hazards for the environment [16,17]. Therefore, evaluating the possible effects of NPs on the organisms used for the biomonitoring of specific contaminants in the contaminated site is needed. Among the various groups of NPs, a wide range of metal-based varieties are common in industry [18]. The effect of NPs on organisms is closely related to their unique characteristics of nano-scale, high surface area, modified surface, and radical formation, especially when considering aggregation of insoluble particles and soluble metals, which greatly affect the state of metal NPs in solution. These characteristics may impact the bioavailability, uptake, and toxicity of metal-based NPs [19].

In this research, comparisons of the effects of NPs were investigated based on the bioluminescence activity of two different recombinant strains, which can be used as bioreporters for the biomonitoring sites. Two different recombinant strains of P. putida mt-2, called KG1206 ($P_m$-lux) and RB1401 ($P_u$-lux), in which expression of the lux gene was placed under the control of a TOL plasmid promoter ($P_m$ or $P_u$), were adopted for investigation. NPs are possible pollutants along with petroleum hydrocarbons, creating a challenging environment for bioreporter analysis.

2. Materials and Methods
2.1. Characteristics of Recombinant Strains

The recombinant strains Pseudomonas putida mt-2 RB1401 and KG1206 contain the intact TOL plasmid and $P_u$-lux and $P_m$-lux fused plasmid, where $P_u$ and $P_m$ are the promoters of the upper and lower operon of the pWW0 (TOL plasmid), respectively. These are responsible for bioluminescence production in the presence of toluene analogs and their intermediates [9,20]. Details of these lux recombinants are shown in Figure 1. Toluene and xylene isomers activate XylR regulatory protein, whereas their intermediates (benzoate and m-toluate), activate XylS regulatory protein. Both inducer-activated XylR and XylS
proteins positively control their promoters $P_u$ and $P_m$, respectively. XylR controls $P_u$ and, also, induces expression of XylS gene, which is responsible for the production of XylS protein as well as regulation of $P_m$. For the KG1206 strain, inducer-activated XylS protein enhances interactions with the promoter $P_m$ to produce bioluminescence from the $P_m$-lux gene fusion in the recombinant strain [9].

![Figure 1. Regulation of recombinant $P_u$-lux and $P_m$-lux genes (●: positive control; ○: regulatory protein) (modified from [9]).](image)

### 2.2. Culture Conditions and Bioluminescence Activity

Tested strains were kept at $-70 \degree C$ using standard procedures [21]. Strains were grown overnight in Luria-Bertani (LB) medium at $27 \degree C$ with shaking (130 rpm). A 1:50 dilution in LB medium was allowed to grow until the optical density was approximately $(\text{OD}_{600}) 0.6$ [9]. An equal mixture of minimum salt medium and a log phase culture grown overnight in Luria-Bertani (LB) medium at $27 \degree C$ with shaking (130 rpm). A 1:50 dilution in LB medium was allowed to grow until the optical density was approximately $(\text{OD}_{600}) 0.6$ [9]. An equal mixture of minimum salt medium and a log phase culture grown in LB was distributed in serum vials (9.9 mL) with inducer (0.1 mL) with inducer (0.1 mL). Vials were sealed with septa to prevent loss of the volatile organic compounds (inducers) [11]. In this experiment, o-chlorotoluene (CT) was used as an inducer with final exposure concentrations of 0.1 to 10 mM. During the incubation periods (generally 5 h), bioluminescence production was measured every 30 min using a Turner 20/20 Luminometer (Sunnyvale, CA, USA), where the maximum detection limit was 9999 RLU (relative light units).

### 2.3. Effects of NPs on Bioluminescence Activity and Metal Analysis

Five tested metal oxide NPs showed following characteristics of size, density, and surface area: CuO (30–50 nm, 6.40 g/cm$^3$, 13.1 m$^2$/g) and NiO (30 nm, 6.67 g/cm$^3$, 50–100 m$^2$/g) were obtained from Nanostructured and Amorphous Materials (Houston, TX, USA); ZnO (40–100 nm, 5.61 g/cm$^3$, 10–25 m$^2$/g), TiO$_2$ (<25 nm, 3.95 g/cm$^3$, 75–85 m$^2$/g), and Al$_2$O$_3$ (40–50 nm, 3.965 g/cm$^3$, 32–40 m$^2$/g) were obtained from Alfa Aesar (Tewksbury, MA, USA). Directly suspended NPs in purified water (pH 7.8) were dispersed for approximately 10 min (40 Hz) by ultrasonic vibration (DH.D250H, DAIHAN, Korea) prior to use. To determine the effects of NPs, the serum vials, which contained culture (0.8 mL) and inducer (0.1 mL o-CT, final concentration 1 mM), were amended with an appropriate concentration of 0.1 mL NP. Concentrations tested for each NP were determined based on a preliminary test as shown in Table S1. The culture was incubated by shaking (130 rpm) at $27 \degree C$ after adding various concentrations of NPs. During the incubation periods (generally 3 h), bioluminescence production was measured in triplicate every 0.5 h. The $EC_{50}$ (effective chemical concentration at which 50% of its effect is observed) values of NPs on bioluminescence activity were estimated using the program SPEARMAN, which is distributed by the US EPA. At the end of the incubation, the solution samples of some experimental sets were filtered (0.45 μm) to measure the concentration of dissolved metal ions using an inductively coupled plasma optical emission spectrometer (Optima 7300DV; Perkin-Elmer Inc., Shelton, CT, USA).
3. Results and Discussion

Previously we reported the bioluminescence activities of two recombinant strains (KG1206 and RB1401) in the presence of toluene, xylene isomers, methylbenzyl alcohol, and their metabolic intermediates (benzoate, m-toluate) under various conditions [9,22]. In this investigation, we compare the effects of NPs on the bioluminescence processes of these two different recombinant strains, which were controlled by different regulatory proteins, while they are induced with o-CT. This inducer had not been examined in our previous investigation, but like toluene, it directly activates the XylR regulatory protein, which controls the Pm promoter positively, resulting in the bioluminescence production from the Pm-lux recombinant gene of the RB1401 strain. This activated XylR which also induces expression from the Ps (XylS promoter) and indirectly activates the XylS regulatory protein, resulting in the production of bioluminescence from the Pm-lux fusion gene of the KG1206 strain [9,20,23]. In addition, a catabolic intermediate of o-CT may also directly activate the XylS regulatory protein, resulting in the production of bioluminescence from the Pm-lux gene of KG1206. A separate report described the transcription of the Pm promoter when XylS is induced by a sufficient concentration of XylR in the presence of direct inducer [24].

3.1. Bioluminescence Activity of Pm- and Pm-lux Gene Fusion to Inducer, o-CT

To examine the changes of bioluminescence activity in KG1206 and RB1401 with time course, cultures were exposed to various concentrations of o-CT ranging from 0 to 10 mM over a period of 5 h (Figure 2). No observable bioluminescence (less than 10 RLU during the exposure period) activity was detected throughout the 5 h of incubation in negative controls. The maximum bioluminescence activity was observed after 2 h incubation for both KG1206 and RB1401, followed by a sharp decrease of bioluminescence activity, and generally bioluminescence activity lasted for approximately 3–4 h. All these activities differ slightly on initial inducer concentration, as was previously published [9]. For example, when KG1206 was exposed to 3 mM of o-CT the maximum bioluminescence was observed at 2 h after induction (2725 ± 461 RLU). For RB1401 this same pattern was seen, although there was a different peak value (7177 ± 654 RLU). These peaks were followed by a steady decline to 70 and 1867 RLU, respectively, after a 3 h incubation (Figure 2a,b).

It is clear from examining a range of inducer concentrations (0.1–10 mM of o-CT) that the peak of bioluminescent activity occurs when 1 mM is used. The peak bioluminescence was achieved after 2 h of incubation, yet the responses were not identical using the two strains. RB1401 produced a maximum of 8783 ± 638 RLU, while KG1206 produced 3617 ± 333 RLU (Figure 2). After maximum bioluminescence activity, the decrease of bioluminescence activity was observed for all exposures and for both strains.
Since α-CT is a direct inducer for RB1401, it was expected that RB1401 would produce greater bioluminescence than KG1206. Depending on the initial inducer concentration, the difference in peak bioluminescence varied greatly. For example, at 1 mM inducer the peak for RB1401 was 2.4 times greater than for KB1206, while the same comparison at 8 mM is 0.72 (Figure 3a). This is also true when the total bioluminescence production is calculated (Figure 3b). The RB1401 strain produced in the range of 4126–31,287 RLU, which was approximately 1.1 to 15 times greater than the bioluminescence compared to that of KG1206, producing 3824–7326 RLU, except at the highly inhibited concentration of 10 mM (Figure 3b). Although the transcriptional components of the regulatory system of the xyl-gene are well described, its architecture is still not clear [25]. Silva-Rocha et al. [25] reported that the action of a metabolic amplifier motif (MAM) is involved in the total regulatory system of the TOL plasmid. MAM appears to express the simultaneous induction of the upper and lower (meta) fragments of the catabolic pathway, which would be difficult to bring about with a standard substrate responsive single promoter. However, based on the results of these preliminary experiments, we selected 1 mM of α-CT as the subsequent experimental conditions.

3.2. Effects of NPs on the Response of Pm-lux Gene Fusion Strain, KG1206

Following the preliminary investigations, various concentrations of individual NP were chosen to investigate the inhibition effects of NPs (ZnO, CuO, NiO, Al2O3, and TiO2) on the bioluminescence production of KG1206 strain with 1 mM of α-CT inducer. The tested concentration ranges for each NP were as follows: CuO 0–40 mg/L, ZnO 0–1 mg/L, NiO 0–2 mg/L, Al2O3, and TiO2 0–5 mg/L (Table S1). The maximum bioluminescence of the control (no NP exposure) during incubation appeared in the range of 1070–2707 RLU after 2 h of incubation.

Effect patterns were slightly different depending on the exposed concentration and NP types, but maximum activity was typically observed after 2 h incubation and the activity lasted for a total of 3 h. Representative of these two NP results are shown in Figure 4. As would be expected, the higher the concentration of NP, the more inhibition of bioluminescence was seen. No stimulation of bioluminescence production was observed under tested concentrations in all cases. For the comparisons of the effects of NPs, all results were presented as the relative values (%) of total bioluminescence produced during 3 h incubation at various concentration ranges of NPs (Figure 5). Total bioluminescence is the sum of bioluminescence measured every 0.5 h from 0.5 h to 3 h of incubation periods. In these representative results, the highest and lowest exposure of each NP showed following activity levels at 2 h incubation: 99% (2056 ± 98 RLU; 1% inhibition) at 10 mg/L CuO, 3.5% (539 ± 166 RLU; 74.1% inhibition) at 40 mg/L CuO, 42% (505 ± 83 RLU; 58% inhibition) at

Figure 3. Comparisons of the bioluminescence activity between KG1206 and RB1401 exposed to various concentrations of α-CT: (a) maximum bioluminescence at each exposed concentration; (b) total bioluminescence production during 5 h incubation periods.
0.2 mg/L Al₂O₃, and 28% (278 ± 108 RLU; 72% inhibition) at 5 mg/L Al₂O₃ (Figure 4). In the case of high concentration Al₂O₃ NP exposure (>1 mg/L), maximum bioluminescence appeared at 1.5 h, slightly earlier than the 2 h of most conditions. Among the NPs tested, ZnO NP showed the highest bioluminescence inhibition effects, showing only 26 ± 4.5% of peak activity at 1 mg/L ZnO (max. exposed concentration). In contrast CuO NP had the lowest inhibition, showing 31 ± 9.7% relative activity at 40 mg/L CuO (maximum exposed concentration) (Figure 5a).

![Figure 4](image)  
**Figure 4.** Effects of CuO and Al₂O₃ NPs on the bioluminescence activity of KG1206 during 3 h incubation period, expressed by the relative activity of the maximum bioluminescence of the control: (a) 0–40 mg/L CuO exposure; (b) 0–3 mg/L Al₂O₃ exposure.

![Figure 5](image)  
**Figure 5.** Comparisons of the changes of relative total bioluminescence with five individual NP (1 mM o-CT inducer): (a) KG1206; (b) RB1401 strains.

### 3.3. Effects of NPs on the Response of Pₐ-lux Gene Fusion Strain, RB1401

The impact of the addition of individual NPs (ZnO, CuO, NiO, Al₂O₃, and TiO₂) on the bioluminescence of RB1401 was also investigated with inducer 1 mM o-CT. The tested concentration ranges for each NPs were slightly different from the set used for KG1206 (CuO 0–100 mg/L, ZnO 0–2 mg/L, NiO 0–3 mg/L, Al₂O₃, and TiO₂ 0–10 mg/L) although the procedure remained the same (Table S1). Representative results for two NPs (CuO and Al₂O₃) are shown in Figure 6, which demonstrate the effect of NP over incubation time. At the highest exposed concentration of 10 mg/L Al₂O₃ and 100 mg/L CuO, the maximum bioluminescence activity appeared at 26% and 36% of the control after 3 h incubation, respectively. Results are presented as the percentage of total possible bioluminescence for...
that strain at 3 h incubation (Figure 5b). No stimulation of bioluminescence was seen with RB1401. However, when compared to the KG1206 results, a slightly different order of the bioluminescence inhibition of exposed NPs appeared in the following order: ZnO > Al2O3 > NiO > TiO2 > CuO. The highest inhibition of bioluminescence production was observed at 2 mg/L ZnO (maximum exposed concentration), showing relative activity 39 ± 1.7%, while the lowest inhibition at 100 mg/L CuO (maximum exposed concentration), showing relative activity 41 ± 2.7% (Figure 5b).

**Figure 6.** Effects of CuO and Al2O3 NPs on the bioluminescence activity of RB1401 during 3 h incubation period, expressed by the relative activity of the maximum bioluminescence of the control: (a) 0–100 mg/L CuO exposure; (b) 0–10 mg/L Al2O3 exposure.

3.4. Comparisons of the Effects of NPs on the Response of *P*~*u*~ and *P*~*m*~-*lux* Gene Fusion Strains

The inhibition differences of individual NPs on the bioluminescence processes of two different recombinant strains were compared using EC50 values, calculated based on the total bioluminescence produced during the exposure period. Of the EC50 values, the inhibition of NPs ranged from 0.25 mg/L (ZnO) to 26.8 mg/L (CuO) for strain KG1206, while the values ranged from 0.42 mg/L (ZnO) to 46.4 mg/L (CuO) for strain RB1401 (Table 1). The toxicity order of the NPs on the bioluminescence activity of KG1206 and RB1401 is nearly identical and was as follows: ZnO (0.25 mg/L) > NiO (0.47 mg/L), Al2O3 (0.68 mg/L) > TiO2 (1.57 mg/L) > CuO (26.8 mg/L) for KG1206, and ZnO (0.42 mg/L) > Al2O3 (1.58 mg/L), NiO (2.92 mg/L) > TiO2 (3.60 mg/L) > CuO (46.4 mg/L) for RB1401. ZnO NP caused the greatest inhibition of bioluminescence activity in both strains, while CuO had the highest EC50s (i.e., the lowest bioluminescence inhibition) for both strains, being in the range of 0.47 mg/L to 1.57 mg/L and 0.42 mg/L to 3.60 mg/L for strain KG1206 and RB1401, respectively. Regardless of the type of NP, the inhibition effects on bioluminescence activity of strain KG1206 were relatively more sensitive than that of RB1401. For both strains, the EC50 values, the inhibition ranked in the order of ZnO > NiO, Al2O3 > TiO2 > CuO for both strains.
Table 1. Summary of the effects of NPs on the bioluminescence activity of two differently recombinated strains.

| Strains   | ZnO NP EC50 (mg/L) | CuO NP EC50 (mg/L) | NiO NP EC50 (mg/L) | Al2O3 NP EC50 (mg/L) | TiO2 NP EC50 (mg/L) |
|-----------|--------------------|--------------------|--------------------|----------------------|--------------------|
| KG1206    | 0.25 ± 0.34 (0.15–0.34) | 26.8 ± 32.47 (22.16–32.47) | 0.47 ± 0.59 (0.38–0.59) | 0.68 ± 0.89 (0.52–0.89) | 1.57 ± 2.52 (0.97–2.52) |
| RB1401    | 0.42 ± 1.54 (0.11–1.54) | 46.4 ± 80.12 (14.35–80.12) | 2.92 ± 4.84 (1.76–4.84) | 1.58 ± 2.22 (1.12–2.22) | 3.60 ± 6.12 (2.12–6.12) |

a values represent mean of triplicates; b values in parentheses represent a 95% confidence level.

The effect of NPs on the activity of various organisms has been extensively investigated in our laboratory, indicating that the toxicity rankings and sensitivities may depend on the organism adopted [26–28] (Table 2). In our previous investigation, most of the organisms tested for NP exposure showed very high EC50 values for TiO2 NPs (>1000 mg/L, 530 mg/L; corresponding to less toxic), but in this study we found very low EC50 values (corresponding to high toxicity) for KG1206 (1.57 mg/L) and RB1401 (3.60 mg/L) (Table 2).

Table 2. Summary of previously published results of the effects of NPs on various biological activities in this laboratory.

| Methods                          | ZnO EC50 (mg/L) | CuO EC50 (mg/L) | NiO EC50 (mg/L) | TiO2 EC50 (mg/L) | Reference |
|----------------------------------|----------------|----------------|----------------|-----------------|-----------|
| Bioluminescence activity         | 1.05 ± 0.54    | 54             | 198            | >1000           | [26]      |
| Seed germination  Lactuca         | 11             | 0.46           | 17             | >1000           | [26]      |
| Seed germination  Raphanus        | 46             | 26             | 114            | >1000           | [26]      |
| ATP contents                     | 11             | 55             | 87             | 530             | [27]      |
| Lactuca root/shoot growth        | 0.50/>2        | 0.25/>1        | 0.57/>5        | ND              | [28]      |

a values represent mean of triplicate; ND, not determined.

Another bioluminescence-producing strain of E. coli previously investigated by this laboratory showed very different toxic effects compared to bioluminescence strains KG1206 and RB1401 used here. EC50 values for bioluminescence producing E. coli were ZnO 1.05 mg/L, CuO 54 mg/L, NiO 198 mg/L, and TiO2 >1000 mg/L, while the two strains examined in this investigation were ZnO 0.25 and 0.42 mg/L, CuO 26.8 and 46.4 mg/L, NiO 0.47 and 2.92 mg/L, and TiO2 1.57 and 4.69 mg/L for KG206 and RB1401, respectively, showing approximately from 5 (ZnO) to over 700 (TiO2) times greater toxicity to bioreporter strains [26] (Table 2). In particular, very significant toxic differences between different bioluminescence-producing strains were observed for TiO2 and NiO NP exposure (Table 2).

More detailed research based on the molecular level is needed to explain the reasons for these significant toxic differences. In this report, the KG1206 strain was found to be more sensitive than RB1401 for all NPs. Other researchers also reported the different effect NPs have with respect to tested organisms. For example, Lin and Xing [29] reported that ZnO NPs tested for seed germination (EC50 range 20–50 mg/L) was less sensitive compared to Daphnia (EC50 range 0.89–1.02 mg/L), which showed very similar sensitivity of the effects on this test [30]. All these results suggested that the appropriate assessment of NPs should be made based on test results of various organisms.

Although the precise toxic mechanisms of NPs on the bioluminescence process are not clear at this point, many studies suggested that NP toxicity on biological systems can be affected by many factors such as solubilized metals, direct contact, characteristics of NPs (i.e., types, shapes, particle size, surface chemistry, residual impurities, etc.) as well as different potential interactions with enzymes involved on specific metabolic processes, and environmental factors [11,30–38]. Factors causing the negative effects might include accumulated...
reactive oxygen species (ROS) in bacterial systems, resulting in damage of membrane and DNA, as well as surface oxidation [39–43]. Studies reported that the production of ROS was induced with ZnO NPs exposure, causing membrane damage and holes in the membrane, leading to cell death by increased membrane permeability [44–46]. Researchers also reported that bacterial enzymes involved in metabolic processes provide numerous sites for NP adsorption and decrease potential interactions with enzymes, generating the significant negative effects on the biosynthetic and catabolic enzymatic activity [35,47]. Therefore, the binding affinities of NPs to various proteins and enzymes produced in the lux-gene metabolic processes can cause differences in toxic effects depending on the type of NPs and recombinant genes of P_m and P_u-lux.

In this investigation, metal concentrations in solution were determined for strain KG1206 experiments to measure the contribution of soluble metals on bioluminescence activity. Dissolved metals were observed as less than 23 µg/L Zn, 77 µg/L Ni, and 648 µg/L Cu, which correspond to 2.3%, 3.8%, and 1.6% of initially amended NPs concentration, respectively. This is a very low concentration. Therefore, the contributions of soluble metals in solution on bioreporter bioluminescence activity were thought to be insignificant or minimal in this investigation. Similar results have been reported, indicating very low solubilized metal ions and no significant concentrations for the total toxicity [48–50]. Bacterial systems are largely protected against NP entry. The particles themselves by the intimate contact between bacteria and NPs could be the main influencing factors on the inhibition effects of NPs, rather than the solubilized metal ions [33,51]. However, once inside the cells, the NPs act as an ion reservoir and were able to dissolve more efficiently, causing the toxicity by increasing ROS production, the oxidization of proteins, and the oxidative DNA damage. However, these could vary depending on the cellular types, biological systems, and test conditions (dose, exposure time, etc.) [50]. Some researchers suggested that complex of NPs and proteins may bind to the cell surface, which enhances cellular uptake and triggers intracellular signaling pathways, or the effect of protein-corona could reduce the interactions with cells or stabilize NPs against solubilization [52–54]. Some possible contaminants of synthesized NPs (e.g., synthesis, breakdown, and surface functionalization) and different conditions of culture and growth may also influence toxicity of NPs regardless of their nano-sized dimensions [35,55,56]. Therefore, the relative contribution of particles and soluble metal ions of metal-based NPs has not yet been clearly described at present and these phenomena may also involve unexpected contributions depending on the test conditions [57,58]. Consequently, focusing an understanding of the interactions between NPs and cellular or molecular mechanism of bioluminescence activity is required for future investigation.

4. Conclusions

The inhibition effects of individual NPs on the bioluminescence activities of P_m-lux (KG1206) and P_u-lux (RB1401) bioreporter fusion strains were compared. The exposure of single NPs showed different inhibition effects depending on the NPs and recombinant strains. In particular, the activity of KG1206 was highly sensitive to the exposure of the NPs compared to the activity of RB1401. Among the NPs tested, ZnO produced the greatest inhibition effect on both strains. These inhibition phenomena can show various results depending on the laboratory test conditions. More practically, environmental contamination by NPs generally exists in mixture state and can also react with soil matrix. Other constituents can also modify their mobility, bioavailability, eco-toxicity, and other properties. Therefore, various methods should be used to examine the real effects of NPs under different conditions in future studies. Although our results showed the wide range of toxic effects of NPs that were dependent on the type of NPs and the endpoints of the tested bioluminescence systems, more systematic and molecular level research is required to clear the long-term and real toxic effects of NPs in environments. Findings in this study provide of great information to build a comprehensive understanding of the potential environmental impacts of NPs. Future studies need to investigate the effects of NPs on
bioluminescent producing processes at the molecular level, and these results will provide clearly additional information about responses on NPs exposure [59].

**Supplementary Materials:** The following are available online at https://www.mdpi.com/article/10.3390/nano11112822/s1, Table S1: Concentration ranges of tested NPs for the study of effects on the bioluminescence activity on two recombinant strains.

**Author Contributions:** K.-S.K. and D.-C.K. designed the experiment and analyzed the data. They also contributed to manuscript preparation. S.L. performed the experiment. I.C.K. performed the research and contributed to writing the paper. R.B. performed Writing—review & editing. All authors have read and agreed to the published version of the manuscript.

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**References**

1. Onwurah, I.N.E.; Ogugua, V.N.; Onyike, N.B.; Ochonogor, A.E.; Otituju, O.F. Crude oil spills in the environment, effects and some innovative clean-up biotechnologies. *Int. J. Environ. Res.* 2007, 1, 307–320. [CrossRef]

2. Varjani, S.J.; Rana, D.P.; Jain, A.K.; Bateja, S.; Upasani, V.N. Synergistic ex-situ biodegradation of crude oil by halotolerant bacterial consortium of indigenous strains isolated from on shore sites of Gujarat, India. *Int. Biodeterior. Biodegrad.* 2015, 103, 116–124. [CrossRef]

3. Ukpaka, C.P.; Lezorghia, S.B.; Nwosu, H. Crude oil degradation in loamy soil using Neem root extracts: An experimental study. *Chem. Int.* 2020, 6, 160–167.

4. Mirjani, M.; Soleimani, M.; Salari, V. Toxicity assessment of total petroleum hydrocarbon in aquatic environments using the bioluminescent bacterium *Aliivibrio fischeri*. *Ecotoxicol. Environ. Saf.* 2021, 207, 111554. [CrossRef] [PubMed]

5. Choi, S.H.; Gu, M.B. Toxicity biomonitoring of degradation byproducts using freeze-dried recombinant bioluminescence bacteria. *Anal. Chim. Acta* 2003, 481, 229–238. [CrossRef]

6. Abbas, M.; Adil, M.; Ehtisham-ul-Haque, S.; Munir, B.; Yameen, M.; Ghaffar, A.; Shar, G.; Tahir, M.A.; Iqbal, M. *Vibrio fischeri* bioluminescence inhibition assay for ecotoxicity assessment: A review. *Sci. Total Environ.* 2018, 26, 1295–1309. [CrossRef]

7. Iqbal, M.; Abbas, M.; Nisar, J.; Nazir, A.; Qamar, A. Bioassays based on higher plants as excellent dosimeters for ecotoxicity monitoring: A review. *Chem. Int.* 2019, 5, 1–80.

8. Saadati, M.; Soleimani, M.; Sadeghsaba, M.; Hemami, M.R. Bioaccumulation of heavy metals (Hg, Cd and Ni) by sentinel crab (*Macrophthalmus depressus*) from sediments of Mousa Bay, Persian Gulf. *Ecotoxicol. Environ. Saf.* 2020, 191, 109986–109992. [CrossRef]

9. Kong, I.C. Effects of binary mixtures of inducers (toluene analogs) and of metals on bioluminescence induction of a recombinant bioreporter strain. *Sensors* 2014, 14, 18993–19006. [CrossRef] [PubMed]

10. Burlage, R.S. Emerging Technologies: Bioreporters, Biosensors, and Microprobes. In *Manual of Environmental Microbiology*, 2nd ed.; Hurst, C.J., Ed.; ASM: Washington, DC, USA, 2002; pp. 147–157.

11. Ko, K.-S.; Kong, I.C. Application of the freeze-dried bioluminescence bioreporter *Pseudomonas putida* mt-2 KG1206 to the biomonitoring of groundwater samples from monitoring wells near gasoline leakage sites. *Appl. Microbiol. Biotechnol.* 2017, 101, 1709–1716. [CrossRef]

12. Auffan, M.; Rose, J.; Bottero, J.Y.; Lowry, G.V.; Jolivet, J.P.; Wiesner, M.R. Towards a definition of inorganic nanoparticles from an environmental, health and safety perspective. *Nat. Nanotechnol.* 2009, 4, 634–641. [CrossRef]

13. Lowry, G.V.; Gregory, K.B.; Apte, S.C.; Lead, J.R. Transformations of nanomaterials in the environment. *Environ. Sci. Technol.* 2012, 46, 6893–6899. [CrossRef]

14. Thuesombat, P.; Hannongbua, S.; Akasit, S.; Chadaichawan, S. Effect of silver nanoparticles on rice (*Oryza sativa* L. cv. KDML 105) seed germination and seedling growth. *Ecotoxicol. Environ. Saf.* 2014, 104, 302–309. [CrossRef] [PubMed]

15. Lines, M.G. Nanomaterials for practical functional uses. *J. Alloys Comp.* 2008, 449, 242–245. [CrossRef]

16. Nowack, B.; Bucheli, T.D. Occurrence, behavior and effects of nanoparticles in the environment. *Environ. Pollut.* 2007, 150, 5–22. [CrossRef] [PubMed]

17. Hänsch, M.; Emmerling, C. Effects of silver nanoparticles on the microbiota and enzyme activity in soil. *J. Plant Nut. Soil Sci.* 2010, 173, 554–558. [CrossRef]

18. Wang, A.; Zhang, L.; Zhao, J.; Xing, B. Environmental processes and toxicity of metallic nanoparticles in aquatic systems as affected by natural organic matter. *Environ. Sci. Nano* 2016, 3, 240–255. [CrossRef]

19. Misra, S.K.; Dybowska, A.; Berhanu, D.; Luoma, S.N.; Valsami, J.E. The complexity of nanoparticle dissolution and its importance in nanotoxicological studies. *Sci. Total Environ.* 2012, 438, 225–232. [CrossRef]
20. Burlage, R.S.; Palumbo, A.V.; Heitzer, A.; Sayler, G.S. Bioluminescent reporter bacteria detect contaminants in soil samples. Appl. Biochem. Biotechnol. 1993, 45/46, 731–740. [CrossRef]

21. Sambrook, J.; Fritsch, E.F.; Maniatis, T. Molecular Cloning: A Laboratory Manual; Cold Spring Harbor Laboratory: Cold Spring Harbor, NY, USA, 1989.

22. Kong, I.C. An optimization of a bioassay for toluene analogs using bioluminescence reporter strain KG1206. Soil Sed. Contam. 2006, 15, 231–239. [CrossRef]

23. Holtel, A.; Marques, S.; Mohler, I.; Jakubzik, U.; Timmis, K.N. Carbon source-dependent inhibition of xyl operon expression of the Pseudomonas putida TOL plasmid. J. Bacteriol. 1994, 176, 1773–1776. [CrossRef]

24. Inouye, S.; Nakazawa, A.; Nakazawa, T. Overproduction of the XylS gene product and activation of the xylDLEGF operon on the TOL plasmid. J. Bacteriol. 1987, 169, 3587–3592. [CrossRef] [PubMed]

25. Silva-Rocha, R.; de Jong, H.; Tamames, J.; de Lorenzo, V. The logic layout of the TOL network of Pseudomonas putida pWW0 plasmid stems from a metabolic amplifier motif (MAM) that optimizes biodegradation of m-xylene. BMC Syst. Biol. 2011, 5, 191. [CrossRef] [PubMed]

26. Ko, K.-S.; Kong, I.C. Toxic effects of nanoparticles on bioluminescence activity, seed germination, and gene mutation. Appl. Microbiol. Biotechnol. 2014, 98, 3295–3303. [CrossRef] [PubMed]

27. Ko, K.-S.; Ha, J.H.; Kong, I.C. Effects of monotypic and binary mixtures of metal oxide nanoparticles on microbial growth in sandy soil collected from artificial recharge sites. Int. J. Mol. Sci. 2015, 16, 27967–27977. [CrossRef] [PubMed]

28. Kong, I.C.; Ko, K.S.; Koh, D.C. Comparisons of the effect of different metal oxide nanoparticles on the root and shoot growth under shaking and non-shaking incubation, different plants, and binary mixture conditions. Nanomaterials 2021, 11, 1653. [CrossRef]

29. Lin, D.; Xing, B. Phytotoxicity of nanoparticles: Inhibition of seed germination and root growth. Environ. Pollut. 2007, 150, 243–250. [CrossRef]

30. Hadijspyrou, S.; Kungolos, A.; Anagnostopoulos, A. Toxicity, bioaccumulation and interactive effects of organotin, cadmium and chromium on Artémia franciscana. Ecotox. Environ. Saf. 2001, 49, 179–186. [CrossRef] [PubMed]

31. Franklin, N.M.; Rogers, N.J.; Apte, S.C.; Batley, G.E.; Gadd, G.E.; Casey, P.S. Comparative toxicity of nanoparticle-catalyzed ZnO, bulk ZnO, and ZnCl2 to a freshwater microalga (Pseudokirchneriella subcapitata): the importance of particle solubility. Environ. Sci. Technol. 2004, 41, 8484–8490. [CrossRef]

32. Horvat, T.; Vidakov, C.Z.; Orescanin, V.; Tkalec, M.; Pevalek, K.B. Toxicity assessment of heavy metal mixtures by Lemma minor L. Sci. Total Environ. 2007, 384, 229–238. [CrossRef]

33. Heinlaan, M.; Ivask, A.; Blinova, I.; Dubourguier, H.C.; Kahru, A. Toxicity of nanosized and bulk ZnO, CuO and TiO2 to bacteria Vibrio Fischeri and crustaceans Daphnia Magna and Thamnocephalus platyurus. Chemosphere 2008, 71, 1308–1316. [CrossRef]

34. Petersen, E.J.; Pinto, R.A.; Landrum, P.F.; Weber, W.J. Influence of carbon nanotubes in pyrene bioaccumulation from contaminated soils by earthworms. Environ. Sci. Technol. 2009, 43, 4181–4187. [CrossRef] [PubMed]

35. MacCorramack, T.J.; Clark, R.J.; Dang, M.K.M.; Ma, G.; Kelly, J.A.; Veinot, J.G.C.; Goss, G.G. Inhibition of enzyme activity by nanomaterials: Potential mechanisms and implications for nanotoxicity testing. Nanotoxicology 2012, 6, 514–525. [CrossRef] [PubMed]

36. Park, J.H.; Lee, K.J.; Cho, J.W.; Jeon, S.L.; Kang, S.H. A study on comparison and analysis of chlorophyll sensor with aceton extraction for chlorophyll measurement in the Nakdong River. J. Kor. Soc. Wat. Wastewat. 2015, 29, 325–335. [CrossRef]

37. Liu, Y.; Baas, J.; Peijnenburg, W.G.M.; Vijver, M.G. Evaluating the combined toxicity of Cu and ZnO nanoparticles: Utility of the concept of additivity and a nested experimental design. Environ. Sci. Technol. 2016, 50, 5328–5337. [CrossRef] [PubMed]

38. Akhdimea, I.D.; Saubadea, F.; Bensona, P.S.; Butlera, J.A.; Olivierb, S.; Kellyb, P.; Verrana, J.; Whiteheada, K.A. The antimicrobial effect of metal substrates on food pathogens. Food Bioprod. Process. 2019, 113, 68–76. [CrossRef]

39. Brayner, R.; Ferrari-Iliou, R.; Brivois, N.; Djediat, S.; Benedetti, M.F.; Fiévet, F. Toxicological impact studies based on Escherichia coli bacteria in ultrafine ZnO nanoparticles colloidal medium. Nano Lett. 2006, 6, 866–870. [CrossRef]

40. Choi, O.; Hu, Z. Size dependent and reactive oxygen species related nanosilver toxicity to nitrifying bacteria. Environ. Sci. Technol. 2008, 42, 4583–4588. [CrossRef]

41. Priester, J.H.; Stoiimenov, P.K.; Mielke, R.E.; Webb, S.M.; Ehrhardt, C.; Zhang, J.P.; Stucky, G.D.; Holden, P.A. Effects of soluble cadmium salts versus CdSe quantum dots on the growth of planktonic Pseudomonas aeruginosa. Environ. Sci. Technol. 2009, 43, 2589–2594. [CrossRef]

42. Gou, N.; Ornis-Hayden, A.; Gu, A.Z. Mechanistic toxicity assessment of nanomaterials by whole-cell-array stress genes expression analysis. Environ. Sci. Technol. 2010, 44, 5964–5970. [CrossRef]

43. Xie, Y.; He, Y.; Irwin, P.L.; Jin, T.; Shi, X. Antibacterial activity and mechanism of action of zinc oxide nanoparticles against Campylobacter jejuni. Appl. Environ. Microbiol. 2011, 77, 2325–2331. [CrossRef] [PubMed]

44. Cohen, R.; Nyska, A. Oxidation of biological systems: Oxidative stress phenomena, antioxidants, redox reactions, and methods for their quantification. Toxicol. Pathol. 2002, 30, 620–650. [CrossRef] [PubMed]

45. Zhang, L.; Jiang, Y.; Ding, Y.; Povey, M.; York, D. Investigation into the antibacterial behavior of suspensions of ZnO nanoparticles (ZnO nanofluids). J. Nanopart. Res. 2007, 9, 479–489. [CrossRef]

46. Dhas, S.P.; Shiny, P.J.; Khan, S.; Mukherjee, A.; Chandrasekaran, N. Toxic behavior of silver and zinc oxide nanoparticles on environmental microorganisms. J. Basic Microbiol. 2014, 54, 916–927. [CrossRef]
47. Peyrot, C.; Wilkinson, K.J.; Desrosiers, M.; Sauvée, S. Effects of silver nanoparticles on soil enzyme activities with and without added organic matter. *Environ. Toxicol. Chem.* 2014, 33, 115–125. [CrossRef]

48. Zhu, X.; Zhu, L.; Duan, Z.; Qi, R.; Li, Y.; Lang, Y. Comparative toxicity of several metal oxide nanoparticle aqueous suspensions to Zebrafish (*Danio rerio*) early developmental stage. *J. Environ. Sci. Health A* 2008, 43, 278–284. [CrossRef]

49. Lee, W.M.; An, Y.J.; Yoon, H.K.; Kweon, H.-S. Toxicity and bioavailability of copper nanoparticles to the terrestrial plants mung bean (*Phaseolus radiatus*) and wheat (*Triticum aestivum*): Plant agar test for water-insoluble nanoparticles. *Environ. Toxicol. Chem.* 2008, 27, 103–112. [CrossRef]

50. Papis, E.; Rossi, F.; Raspanti, M.; Dalle-Donne, I.; Colombo, G.; Milzani, A.; Bernardini, G.; Gornati, R. Engineered cobalt oxide nanoparticles readily enter cells. *Toxicol. Lett.* 2009, 189, 253–259. [CrossRef]

51. Hu, C.W.; Li, M.; Cui, Y.B.; Chen, J.; Yang, L.Y. Toxicological effects of TiO$_2$ and ZnO nanoparticles in soil on earthworm *Eisenia fetida*. *Soil Biol. Biochem.* 2010, 42, 586–591. [CrossRef]

52. Chattopadhyay, S.; Dash, S.K.; Tripathy, S.; Das, B.; Mandal, D.; Pramanik, P.; Roy, S. Toxicity of cobalt oxide nanoparticles to normal cells; an in vitro and in vivo study. *Chemico-Biol. Interact.* 2015, 226, 58–71. [CrossRef]

53. Lu, B.; Smith, T.; Schmid, J.J. Nanoparticle-lipid bilayer interactions studies with lipid bilayer arrays. *Nanoscale* 2015, 7, 7858–7866. [CrossRef] [PubMed]

54. Bossi, E.; Zanella, D.; Gornati, R.; Nernardini, G. Cobalt oxide nanoparticles can enter inside the cells by crossing plasma membranes. *Sci. Rep.* 2016, 6, 22254. [CrossRef] [PubMed]

55. Zoroddu, M.A.; Medici, S.; Ledda, A.; Nurchi, V.M.; Lachowicz, J.I.; Peana, M. Toxicity of nanoparticles. *Curr. Med. Chem.* 2014, 21, 3837–3853. [CrossRef] [PubMed]

56. Wu, F.; Harper, B.J.; Harper, S.L. Differential dissolution and toxicity of surface functionalized silver nanoparticles in small-scale microcosms: Impacts of community complexity. *Environ. Sci. Nano* 2017, 4, 359–372. [CrossRef]

57. Al-Bairuty, G.A.; Boyle, D.; Henry, T.B.; Handy, R.D. Sublethal effects of copper sulphate compared to copper nanoparticles in rainbow trout (*Oncorhynchus mykiss*) at low pH: Physiology and metal accumulation. *Aquat. Toxicol.* 2016, 174, 188–198. [CrossRef]

58. Razmara, P.; Lari, E.; Mohaddes, E.; Zhang, Y.G.; Goss, G.; Pyle, G.G. The effect of copper nanoparticles on olfaction in rainbow trout (*Oncorhynchus mykiss*). *Environ. Sci. Nano* 2019, 6, 2094–2104. [CrossRef]

59. Vandenbrouck, T.; Soetaert, A.; van der Ven, K.; Blust, R.; de Coem, W. Nickel and binary metal mixture responses in *Daphnia magna*: Molecular fingerprints and (sub)organismal effects. *Aquat. Toxicol.* 2009, 92, 18–29. [CrossRef] [PubMed]