Assessment of Land Degradation in Semiarid Tanzania—Using Multiscale Remote Sensing Datasets to Support Sustainable Development Goal 15.3

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Abstract: Monitoring land degradation (LD) to improve the measurement of the sustainable development goal (SDG) 15.3.1 indicator (“proportion of land that is degraded over a total land area”) is key to ensure a more sustainable future. Current frameworks rely on default medium-resolution remote sensing datasets available to assess LD and cannot identify subtle changes at the sub-national scale. This study is the first to adapt local datasets in interplay with high-resolution imagery to monitor the extent of LD in the semiarid Kiteto and Kongwa (KK) districts of Tanzania from 2000–2019. It incorporates freely available datasets such as Landsat time series and customized land cover and uses open-source software and cloud-computing. Further, we compared our results of the LD assessment based on the adopted high-resolution data and methodology (AM) with the default medium-resolution data and methodology (DM) suggested by the United Nations Convention to Combat Desertification. According to AM, 16% of the area in KK districts was degraded during 2000–2015, whereas DM revealed total LD on 70% of the area. Furthermore, based on the AM, overall, 27% of the land was degraded from 2000–2019. To achieve LD neutrality until 2030, spatial planning should focus on hotspot areas and implement sustainable land management practices based on these fine resolution results.

Keywords: land degradation neutrality; SDG; land productivity; land cover; NDVI; Landsat; vegetation-precipitation relationship; soil organic carbon; Google Earth Engine

1. Introduction

Land degradation (LD) is defined as the “continuous reduction or loss of the productivity of the land due to a combination of natural and anthropogenic causes” [1]. It is a global problem and affects people, their livelihoods and nature. Studies suggest that up to 3.2 billion people live and depend on degraded lands [2] and that approximately a quarter of the world’s lands are affected by LD [3,4]. Poor people, who often rely on agriculture, are most vulnerable to LD [5,6]. Lost ecosystem services due to land use and land cover (LULC) change and LD account for up to USD 10.5 trillion loss per year, which is about a sixth of the world’s gross domestic product (GDP) [7]. Furthermore, biodiversity is declining globally, with tremendous losses in sub-Saharan Africa because of LD [6]. Projections suggest that lower productivity in the face of climate change will drive LULC change globally. Moreover, the population growth, combined with a changing diet, will have an enormous influence on agriculture and thus LD [8]. It is for these reasons that the world community introduced the sustainable development goal (SDG) 15.3, which aims to
“restore degraded land and strive to achieve an LD-neutral world”, highlighting the global importance of this issue [9,10].

Tanzania is a hot spot of LD, with more than half its area showing signs of degradation [2,11]. It has the highest annual forest area net loss in East Africa and the fifth-highest worldwide [12]. The cost of LD has been summed up to USD 2.3 billion annually in the first decade of the new millennium [13]. Seventy-five percent of the total labor force, mostly rural people, work and depend on the agricultural sector, which is accountable for about 30% of the GDP [14]. Although the cultivated area increased in the last years, the output per hectare (ha) decreased, both in annual and perennial crops, even though fertilizer consumption quadrupled at the same time [15]. The number of undernourished people is growing and is currently more than 30% [16]. The population is increasing while agricultural productivity is stagnating, and the economic dependency on natural goods is still high. The consequences of this dilemma area persisting pressure on land and, thus, a probable conversion of natural into cultivated land in the coming years. The poor people’s food security is also at risk, and in the coming years, in the face of climate change, new insecurities are likely to arise [17]. This holds especially true for the rural semiarid central districts of Kiteto and Kongwa (KK).

Agricultural intensification and sustainable land management (SLM) are keys to halt and reverse LD [18–20]. One major constraint that prevents action is the lack of spatial information on the extent and magnitude of LD [18]. In contrast to the laborious fieldwork, remote sensing offers the unique opportunity to consistently assess vast areas over a long period [2–4]. Unfortunately, the existing LD maps have a coarse spatial resolution and provide inconsistent estimates of the affected area [8]. For example, previous estimates of the extent of LD in Tanzania range from 41% to half of the country [2,3,11]. These variations emanate from differences in definitions of LD, monitoring methods and lack of appropriate data [6,21]. In the course of SDG 15.3 implementation, standard methods for assessing LD were introduced, making reports more comparable.

This new standard methodology, recommended by the United Nations Convention to Combat Desertification (UNCCD), includes the usage of three sub-indicators for the complimentary assessment of LD [22]. The first sub-indicator, land cover (LC), reports changes in vegetation cover. The second, land productivity (LP), captures changes in ecosystem functions. The last, soil organic carbon (SOC), indicates slower changes resulting from biomass alterations [20]. The three sub-indicators are aggregated to form the land degradation indicator. Improvements in one indicator cannot compensate losses in others, as they are complementary and not additive. Thus, the “one-out, all-out” approach is applied whereby even if one indicator shows signs of decline and the others are positive, the land is deemed to be degraded [23].

The recent Tanzanian national LD-neutrality (LDN) report follows these guidelines [24]. However, it only assesses LD for the first ten years of the 21st century and mainly uses global default data with a coarse spatial resolution. The 1 km coarseresolution is inadequate to monitor LD in small mountainous and highly fragmented landscapes, as it may miss out on smaller than pixel size LD areas [25].

Overall, only a few studies have been published on the subject of SDG 15.3.1 monitoring and assessment. Gichenje and Godinho [26], for example, conducted a baseline assessment of the SDG indicator 15.3.1 for the years 1992 to 2015 using the Advanced Very High Resolution Radiometer (8 km, AVHRR) Normalized Difference Vegetation Index (NDVI) time series and the European Space Agency (ESA) Climate Change Initiative (CCI) LC map in Kenya. In Mozambique, Frederique et al. [27] analyzed only the LD sub-indicator LP trend using the Moderate Resolution Imaging Spectroradiometer (250 m, MODIS) NDVI from 2001 to 2016.

However, these studies share the common disadvantage of applying only default methodology and global datasets for national and subnational LD assessments. Though Akinyemi et al. (2020) used a customized 30m resolution LC map to assess the LC sub-indicator of SDG 15.3.1 in Botswana, this study relied on AVHRR time-series assessment
for the LP sub-indicator. Furthermore, no studies exist in Africa that used high-spatial-resolution datasets the assessment of more than one sub-indicator of SDG 15.3.1. Therefore, it is vital to overcome the existing research gaps and use high-resolution spatial data to provide improved information on the SDG 15.3 [28].

In this light, the main aim of our study was to assess the SDG 15.3.1 indicator based on the newly adopted approach based on the higher resolution (compared to default UNCCD datasets) 30m Landsat time series and 30m LC maps and compare our results to the estimates of the SDG 15.3.1 based on the default UNCCD data and methods.

Our study addressed the following research questions:

- How much land is degraded, and where are the hotspots of LD in KK?
- How do the individual sub-indicators affect LD?
- Does using higher resolution data (30 m) improve the delineation of LD compared to moderate-resolution data (250 m)?

2. Materials and Methods

2.1. Study Area

The study site is situated in Kiteto and Kongwa districts, located in Dodoma and Manyara regions of Central Tanzania, respectively (Figure 1). The elevation ranges between 850 and 2100 m above sea level. The study area has a hot arid steppe climate [29]. The average monthly temperature stays between 19 and 25 °C all year, and the precipitation is roughly 600 mm a year, with interannual differences of 500 to 800 mm. Large parts of northern Kiteto and more minor areas of the mountainous region in Kongwa are protected areas for nature and landscape conservation.

2.2. Materials

The SDG 15.3.1 indicator and its three LDN sub-indicators were computed using the recommended default method (DM) with Trends.Earth [30] and the adapted methods (AM) using high-resolution Landsat (and other) datasets (Table 1).

Figure 1. Location of the study area in Central Tanzania (A,B) and protected areas (C).
The DM (LC) map provided by the UNCCD is based on the 300 m ESA CCI LC map (Table 1). The AM utilized 30 m LC maps for 2000–2018 in the study area that the Regional Centre for Mapping of Resource for Development (RCMRD) developed. Both datasets were disaggregated into the six LC classes as defined by Intergovernmental Panel on Climate Change (IPCC), i.e., forestland, grassland, cropland, wetland, urban, and otherlands [31].

The recommended global default dataset uses the MOD-13Q1-coll6 (250 m) MODIS-NDVI products [30]. In contrast, the AM was calculated based on a 30 m resolution NDVI from a combination of Landsat 5, 7 and 8 (Table 1). The Landsat time series were accessed and analyzed using Google Earth Engine [32], based on atmospherically corrected surface reflectance collections (Table 1). The Landsat 5 and 7 data were spectrally harmonized with Landsat 8 series using linear transformation [33]. As a further step to improve the image quality, the fmask was adopted to mask out clouds and cloud shadows [34,35]. Generally, the images with cloud cover scores higher than 80% were removed. Finally, the NDVI was calculated for each image, and then the images of the same admission time were merged and clipped to the extent of the study area. As it is recommended to constrain the observation period to the growing season to reduce the number of irrelevant assets for the computation and enhance the quality of the time series [22], we used the imagery from November to June. When using Trends.Earth, there is no possibility to apply the computation to the growing season, so the DM uses the whole calendar year. In order to integrate the rainfall information, data from Climate Hazards Group InfraRed Precipitation with Station (CHIRPS) were used (Table 1).

**Table 1.** The datasets used for land degradation neutrality (LDN) reporting. For the land cover, the European Space Agency Climate Change Initiative (ESA-CCI) and the Regional Center for Mapping of Resources for Development (RCMRD) were used. Land productivity is based on the Moderate Resolution Imaging Spectroradiometer (MODIS) and Landsat. The precipitation is based on the Climate Hazards Group InfraRed Precipitation with Station data (CHIRPS). Lastly, the soil organic carbon content is derived from SoilGrids250m.

| LDN Sub-Indicators | Method | Data | Resolution/Year | Reference |
|--------------------|--------|------|------------------|-----------|
| Land Cover         | Default Method (DM) | ESA-CCI | 300 m (2000–2015) | [36] |
|                    | Adapted Method (AM) | RCMRD | 30 m (2000–2018) | [37] |
| Land Productivity  | DM     | MOD-13Q1-coll6 | 250 m (2000–2015) | [38] |
|                    | AM     | Landsat 5     | 30 m (2000–2013) | [39] |
|                    |        | Landsat 7     | 30 m (2000–2019) | [39] |
|                    |        | Landsat 8     | 30 m (2013–2019) | [39] |
|                    | DM/AM  | CHIRPS        | 0.05 arc° (2000–2019) | [40] |
| Soil Organic Carbon| DM/AM  | SoilGrids250m | 250 m          | [41] |

The SOC metrics were derived from the SoilGrids250m dataset [41] for the DM and the AM, as there is no national SOC database for Tanzania. SOC is measured at a depth of 30 cm and is stated as mass per area (e.g., tons per hectare (t/ha)) [22].

### 2.3. Methods

The calculation of the SDG 15.3.1 indicator is based on the “one out, all-out” approach (Ref. [23] and Figure 2). The three LD sub-indicators (LC change, LP decline and loss of SOC) are estimated, and if one indicator signals degradation, the LD indicator will reflect this as well. A baseline is needed to compare the progress of LDN. The baseline year (t₀) was set to be 2015 and is computed as the average of the period leading up to t₀ (2000–2015). The indicators are then remeasured in regular time intervals leading to 2030, and change is used to monitor the progress to accomplish LDN [20].
To calculate the indicator for the reporting year 2019 ($t_1$), it is necessary first to assess the baseline until $t_0$ and then calculate the change from the baseline to $t_1$ (Figure 2). As a final step, combine both results. The details of the calculation of each indicator are explained in the following section. The three LD sub-indicators were created from satellite images using cloud-based geospatial computing. The indicators were calculated using Trends.Earth [30] and Google Earth Engine [32] for the DM and AM, respectively. As Trends.Earth currently only enables the computation for the baseline period (BP), the DM is only available from 2000 to 2015.

2.3.1. Sub-Indicator 1: Land Cover Transitions and Degradation

The first SDG 15.3.1 indicator is the LC change. To assess the LC degradation, the transitions between 2000–2015 and 2015–2018 were analyzed for the baseline and the first monitoring period (MP), respectively. To determine whether changes from one LC class to another are interpreted as degradation, a change matrix can help visualize the transitions (Table 2) based on the Good Practice Guidance by the UNCCD [31]. It is recommended to adopt this matrix for the national context. Therefore, transitions from grasslands to croplands were not considered LD for the AM to avoid tradeoff between ecosystems and food security and between nomadic and sedentary living.

2.3.2. Sub-Indicator 2: Loss of Land Productivity

LP is described as “the biological productive capacity of the land”. It is closely associated with net primary productivity [42], which can be measured directly with earth observation methods [22]. NDVI is a widely used index detecting LP [26,43,44]. The LP sub-indicator consists of three distinct components, namely trend, state and performance.

The LP trend component measures the trajectory of change in productivity over time. It is calculated at the pixel level using linear regression and the Mann Kendall significance test [22,45,46]. Positive and negative changes in NDVI indicate increasing and decreasing productivity associated with vegetation recovery and degradation, respectively. The eight most recent years of data were used to create a new distinct and significant time series that is more responsive to present land conditions. Further, following [47], we accounted
for the effect of rainfall variability on vegetation productivity trends by using the rain use efficiency (RUE) method.

The LP state component represents recent changes in LP compared to the BP. The yearly NDVI mean images of the shortened BP (2000–2012) were normalized and assigned to classes from 1 to 10 based on their percentiles. To avoid annual fluctuations, contemporaneous values of the three-year antecedent \( t_0 \) and \( t_1 \) were classified in this scheme. Areas with a reduction of two or more classes were classified as degraded, while the rise by two categories was interpreted as an improvement [31].

The LP performance component examines local productivity compared to similar ecoregions defined by the unique combination of SoilGrids [41], soil taxonomy great groups and LC classes (Table 1). The 90th percentile in each ecoregion was calculated as a proxy for the maximum productivity level. The LP performance was then calculated based on the ratio of the observed mean NDVI value per pixel and the NDVI_{max(90th)}. Values below 0.5 indicate regions where the LP is low and LD may prevail [31].

The overall LP sub-indicator is calculated based on the three components mentioned earlier. As the LP trend is based on a statistically significant test, it is most influential, and its status determines LP degradation. Only if both LP status and LP performance show negative results, does the LP indicator also show degradation [22]. If only the LP state component shows degradation, this could indicate “early signs of decline” because the other indicators may not have detected the most recent LD. Further, if only performance shows degradation, there is no temporal trend, and the land is classified as “stable but stressed” [22]. In contrast to the Good Practice Guidance by UNCCD, Trends.Earth (DM) also incorporates the “early signs of decline” state component into the LP degradation [30].

| AM Land Cover Category in 2015 (km²) | 2000 Total (km²) |
|-------------------------------------|-------------------|
| **Forestland**                      | 2519              |
| **Grassland**                       | 8094.6            |
| **Cropland**                        | 3969.3            |
| **Wetland**                         | 297.1             |
| **Urban**                           | 193.8             |
| **Otherland**                       | 2017.5            |

2.3.3. Sub-Indicator 3: Degradation of Soil Organic Carbon

The Good Practice Guidance for the SOC sub-indicator is based on the maximum equilibrium SOC content at a location that is controlled by environmental factors such as rainfall, evaporation, solar radiation, and temperature [22]. The content can change based on three distinct change factors: First, the land-use factor represents SOC stock changes based on the type of land use. Second, the management factor reflects the management practice of the land use (e.g., grazing intensity on grasslands). Third, the input factor represents the different amounts of carbon input into the soil [22,48,49]. While the LULC change factor can be used with LC as a proxy, there are presently no sufficient datasets
available to provide information about the management or the input for the other two indicators. Thus, the only indicator to assess SOC changes is the second LD indicator LC change [22].

3. Results

Three sub-indicators, namely LC transitions, LP decline, and SOC loss, were estimated to derive the SDG 15.3.1 indicator using the default and adapted methods. The patterns of each sub-indicator based on DM and AM are described in the following sections starting with the BP from 200 to 2015 for both DM and AM. The first monitoring period from 2015 to 2019 is only assessed using the AM, as the data necessary for this period are currently not available in Trends.Earth.

3.1. Sub-Indicator 1: Land Cover Transitions and Degradation

According to the DM based on the medium-resolution 300 m LC maps, over 99% of the study area remained stable in the BP (2000–2015) (Table A1). Urban areas covering less than 0.1% of the study area experienced the highest relative expansion (56%). The forestlands were the only other LC class that increased in area significantly (4.4%) in the BP.

In contrast to the DM, the AM with high-resolution (30 m) LC data revealed that 6.7% of the total area changed to a less desirable LC class, signifying LD, and only 2.3% of analyzed areas improved. The area of (semi)natural LC, such as forestlands (−19%), grasslands (−6.6%) and wetlands (0.1%), mostly declined, whereas the croplands recorded the highest spatial gain (24.2%) (Figure 3 and Table 2).

Figure 3. Sankey plot describing the land cover transitions between the years 2000, 2015 and 2018 using high-resolution land cover data. Bands represent the actual proportion of land that changed class over time.
The trend observed in the BP continued in the first years of the MP (Figure 3 and Table A2). Overall, from 2015 to 2018, 3.3% of the total area was degraded during the MP, while 1.2% of the area changed to a more desirable LC. Grass- and forestlands continued to decline by 3 to 9%, respectively, while anthropogenic(-influenced) covers such as cropland and urban areas expanded. Compared to about 3000 ha forests lost per year (a) in the BP, the rate doubled to 6000 ha/a in the MP. Similarly, the changes in croplands increased from 6000 ha/a in 2000–2015 to 7500 ha/a in 2015–2018.

3.2. Sub-Indicator 2: Loss of Land Productivity

The DM revealed that the LP sub-indicator showed degradation in 71.1% of the area during the BP from 2000 to 2015 (Table 3). The LP component trend showed “decline” in 26.8% of the area (Figure A1A). Another 44.3% of the study area showed “early signs of decline” (LP component state, Figure A2A), and the rest (28.9%) remained stable (Figure 4A). According to DM, croplands were most affected (48.4%) by LP decline in 2000–2015 (Figure 5). Forestlands with only about 11.7% marked as degraded were less affected compared to their actual LC share (Figure 5).

Table 3. The land productivity (LP) status in percent for the default (DM) and adapted methods (AM) for the baseline period from 2000 to 2015, as well as for the first monitoring period of 2015-2019. Furthermore, the land cover share of the degraded area in the target year is depicted.

| LP Status (%) | DM 2000–2015 | AM 2000–2015 | AM 2015–2019 |
|---------------|--------------|--------------|--------------|
| Degraded      | 71.1         | 8.2          | 12.2         |
| Stable        | 28.9         | 91.3         | 87.7         |
| Improved      | 0            | 0.5          | 0.1          |

Figure 4. The land productivity sub-indicator generated using (A) the default approach with MODIS imagery, (B) the adapted approach with Landsat imagery for the baseline period, and (C) the adapted approach with Landsat imagery for the monitoring period 2015–2019.
Figure 5. The bar chart showing the distribution of land productivity (LP) decline sub-indicators over the land cover classes using the default (DM) and adapted methods (AM). The dashed lines show the actual land cover share.

Based on the AM applied between 2000 and 2015, the final composite indicators of LP decline revealed that 8.2% of the study area was degraded between 2000 and 2015 (Table 3). This is nearly entirely based on the 8.2% “decline” of the LP trend component (Figure A1B). Further, 9.1% and 1.4% of the study area were marked as showing “early signs of decline” (Figure A2B) and “stable but stressed” areas (Figure A3B), respectively (Figure 4B). Grass- and croplands accounted for 43.5% and 42% of the degraded area (Figure 5). The decline in forestlands was, in turn, detected only on 2.6% of the total degraded area.

LP declined over 12.2% of the study area during the MP from 2015 to 2019 (Figure 4C). With an increase from 9.1% up to 17% of the area, the share of areas with “early signs of decline” (state component) was higher than during the BP (Figure A2). The area where LP was improving was reduced from 855 to 171 km$^2$ compared to the BP.

3.3. Sub-Indicator 3: Degradation of Soil Organic Carbon

Soil organic carbon was not directly computed but rather assessed through LC classes’ alteration and the related change factors [49]. SOC did not change significantly with the DM during the BP from 2000 to 2015: 99.9% of the land did not change the in SOC content by more than 10% (Table 4). Changes in the individual LC classes were also neglectable.

In contrast to DM, the AM approach revealed that during the BP of 2000–2015, 8.4% of the land was degraded due to SOC diminishment, while 2.1% increased in SOC content (Figure 6). The average SOC stock declined from 51.2 to 50.2 t/ha in 2015, losing 1,592,423 t of carbon over 16 years (Table 4). Forestlands had significantly higher SOC stocks (62.2 t/ha) at $t_0$ than the other LC classes. Based on the transitions in LC, the amount of SOC in forests dropped by 19%, while SOC under agricultural use increased by 25.1%.
In the MP, the SOC content experienced significant losses on 3.7% of the land. The same trend was observed in other LC classes (forest-, grass- and wetlands) that gradually lost SOC in the MP (Table 4).

Table 4. The soil organic carbon (SOC) content for the default (DM) and adapted methods (AM) for the baseline period from 2000 to 2015 as well as for the first monitoring period of 2015–2018.

| Status (%) | DM SOC | AM SOC | AM SOC |
|------------|--------|--------|--------|
|            | 2000   | 2015   | 2000   | 2015   | 2018   |
| Degraded   | 0.1    | 8.1    | 3.7    |
| Stable     | 99.9   | 90     | 94.7   |
| Improved   | 0      | 2      | 1.7    |

| SOC (t/ha) | Study area | Forestland | Grassland | Cropland | Wetland | Urban | Otherland |
|------------|------------|------------|-----------|----------|---------|-------|-----------|
|            | 51.2       | 54.7       | 55        | 46.2     | 45.1    | 36.2  | 0         |
|            | 51.2       | 54.7       | 55        | 46.2     | 45.1    | 36.2  | 0         |
|            | 51.2       | 63.2       | 50.7      | 46.5     | 49.2    | 39.5  | 46.2      |
|            | 50.2       | 62.2       | 49.7      | 46.9     | 47      | 42.8  | 47.6      |
|            | 49.9       | 62         | 49.5      | 46.9     | 46.7    | 42.8  | 47.6      |
3.4. Combined Sustainable Development Indicator 15.3.1 for the Baseline and First Monitoring Period

During the BP from 2000 to 2015, the DM method identified 71.1% of KK’s area as degraded and only as 0.5% improved (Figure 7A). This result is mainly caused by the sub-indicator LP, while the two other indicators LC and SOC showed nearly no degradation. The LP degradation was mainly driven by the state component of LP in 70.3% of the total area.

On the contrary, during the BP, the AM showed that 16.4% of the area was degraded and 2.7% improved (Figure 7B). The distinct sub-indicators influenced the final indicator more evenly with 52.4%, 50% and 31.7% by SOC, LP and LC, respectively, compared to the DM.

The AM for the first MP (2015–2019) showed that 16% of the total area was degraded, 1.5% improved and more than 82% remained stable (Figure 7C). Forests and grasslands were the least affected among LC classes. Croplands (38%) and wetlands (7%) experienced the most degradation between 2015 and 2019. Over three-fourths of the degradation was driven by the LP sub-indicator, whereas LC and SOC only contributed 20% and 23% to LD, respectively.

3.5. Combined Sustainable Development Indicator 15.3.1 over 20 Years Using the AM

Over the whole period of 20 years (2000–2019), which results in the SDG 15.3.1 indicator at timestep $t_1$, 27.7% of KK was degraded, and 2.8% of KK improved (Figure 8A). Thus, the LD was widespread across the two studied districts and formed several LD clusters (Figure 7B,C). The degradation was not equally distributed over the study area: the biggest LD hotspots were Central and Western Kiteto, as well as Western Kongwa (Figure 8A). Even though the land covered by forests decreased and the land covered by crops increased from 2015 to 2018, the degraded proportion changed conversely as follows: The degraded area covered by forests increased to 3.9%, while the area covered by crops sank to 41.9%. While SOC’s degraded area only changed slightly, the relative contribution sank from 50 to 30% (Figure 8B). The degraded area, which is solely influenced by LP, rose over 50% and interplayed with others over 70%.
4. Discussion

The presented study is the first in Africa to support the monitoring of the SDG 15.3.1 indicator using fine-spatial-resolution (30 m) satellite time series data for LD assessment. This is a key contribution considering that previous studies used 250 m to 8 km resolution data [24,26,50] for LP sub-indicator monitoring, unlike our study that utilized long-term Landsat time series for SDG 15.3.1 monitoring. Furthermore, it is the first sub-national study that assesses the SDG 15.3.1 indicator in Tanzania for the BP and includes the MP until 2019. The first 4 out of 15 years of the SDG time frame are assessed and could help identify hotspot areas for targeting the appropriate measures to combat LD in the study area.

The presented LD assessment in KK districts confirmed that the LD problem is acute in Tanzania. The Tanzanian target is to achieve LDN by 2030 [24]. Both KK are part of declared LD hotspot regions, which need to improve 25% of the area based on the status at t₀. According to our analysis, only 2.7% of the land area has improved and 27.7% is degraded. Next to the (sub)national targets, there are also specific targets to avoid, minimize and reverse LD in Tanzania [24]. Among others, about half of the current national forest area should be restored, 50% of the national croplands should improve LP and the SOC content in croplands should rise to 54.5 t/ha [51]. Despite these more specific and ambitious targets, our results show a negative trend in all LD sub-indicators analyzed, suggesting that more efforts are needed to combat LD in the study area.
Precisely, instead of restoring forest areas, even more trees were cut over 19 years (14.7% to 10.9% tree cover). In croplands, LP degradation was above average, while the SOC content in croplands improved marginally. A possible explanation could be that restoration attempts using SLM practices had not yet shown effects, because it takes several years for the change to be monitored remotely \[52,53\]. Moreover, it takes decades for SOC to change \[49,54\]. Hence, it is of paramount importance to prioritize the detected LD hotspots for rehabilitation and SLM practices to reverse LD processes.

There are currently no sub-national studies for KK districts. With around 27% of the area in KK being degraded, it is less affected by LD compared to national assessments found in \[2,3\] or \[11\]. However, the comparison with these studies is difficult, as they used different monitoring periods (ending in the 2000s and 2016) and only a subset of the methodology (LP trend) and coarse resolution imagery (i.e., 8 km AVHRR data). This suggests that our study brought LD assessment in Tanzania one step further by assessing three components of LD according to the SDG 15.3.1 indicator. Further, using significantly higher spatial resolution, spatial datasets allowed us to reveal spatial patterns of LD beyond pixel sizes of 8 km \[2,3,11\] or 1 km \[24\].

Our study compared the results of the LD assessments based on default UNCCD-suggested datasets (250 m MODIS data used for LP sub-indicator and 300 m ESA CCI LC maps) and customized relatively high-resolution datasets (30 m Landsat data used for the LP sub-indicator and 30 m RCMRD LC maps). The resulting differences between LD estimates based on DM and AM were striking and could be primarily attributed to the difference in the pixel size of 6.25 ha (MODIS) versus 0.09 ha (Landsat), which could be critical in specific areas where fine LD patterns prevailed. This finding is confirmed by several studies highlighting the importance of using high-resolution imagery to detect LD, especially on heterogeneous landscapes, such as KK districts, dominated by heterogenous small-scale farms \[50,55,56\]. Recent studies that used ground-truth data for validation showed that using Landsat data for the LC sub-indicator captured LD better than using ESA-based 300 m datasets \[50\]. Nevertheless, certain factors could have impacted the AM, such as the scan-line failure in Landsat ETM+ data. To reduce the potential negative influence of this on our analysis, we applied several preprocessing steps confirmed to be effective in similar studies \[56\].

NDVI was applied in this study, although it was affected by soil brightness in areas with low vegetation cover. Other vegetation indices, such as MSAVI or MSAVI2, are less sensitive to soil optical properties in less vegetated areas and, therefore, can be used to detect a decline in vegetation productivity \[57\]. However, the alternative indices have significantly better results than NDVI only in areas where bare soils prevail. Further, Tüshaus et al. \[58\] compared NDVI with the Soil-Adjusted Vegetation Index (SAVI) and MERIS-based Terrestrial Chlorophyll Index (MTCI). The results indicated only little differences between the different vegetation indices. Nevertheless, the impact of different vegetation indices on the estimated LDN sub-indicators can be further tested.

Furthermore, our results pointed out that the ESA CCI LC did not reflect significant LC changes during the BP in KK districts. Other local estimates, such as the National Forest Resources Monitoring and Assessment of Tanzania Mainland \[21\] or Tanzanian Forest Reference Emission Level \[59\], suggest a change rate that is three to twenty times higher, respectively, for a similar period analyzed. Our result is in line with the study of Kimaro et al. \[60\], who investigated the LC change for the study area from 1987 to 2010. Their study indicated that the LC change was already in progress over 30 years ago with heavy declines in (semi)-natural landscapes. This suggests that our research offers advancement of sub-national assessment of LD in heterogeneous landscapes.

Our study revealed that the LP sub-indicator impacted LD in the study area the most (by 50%) using the AM. The remaining half is affected by SOC, LC, or by the combination of more than one sub-indicator. On the other hand, the LD indicator using the DM is nearly solely affected by the LP sub-indicator, which is primarily driven by the state component. This suggests two things: First, our AM is better suited to reflect the ongoing
multidimensional degradation in KK districts. Second, even if the ongoing LULC change stops, the degradation will not halt because of the decline in LP.

This is well reflected in croplands, which were the worst affected land cover class, not only in LP decline but also in SOC loss. Due to the continuous cultivation of the agricultural lands combined with overgrazing and little fertilizer inputs, the crop yields in the study area are reportedly low, caused by the limited availability of soil nutrients and organic matter content [18]. Another study that assessed LD in Kenya in similar environmental and land use settings found that croplands experienced the highest decline in LP, indicating that unsustainable farming practices are widespread throughout Eastern Africa [26]. This has serious consequences, as already 30% of the Tanzanian population are undernourished [16], and the yield gap for the main crops needs to be closed for the population to sustain itself in the coming decades [61].

The soils in KK districts lost 1.6 million t of SOC due to LULC change from 2000–2018, according to our study. This is especially dire, as SOC is vital for soil quality and is a key ecosystem indicator [62]. The study by van der Esch et al. [63] suggests that due to LULC change, 27 Gt of SOC will be further lost globally by 2050, mainly in sub-Saharan Africa. Studies conducted in Tanzania found that higher SOC values on the farm level resulted in financial benefits for the farmers [64]. Thus, increasing SOC via SLM practices would not only improve farmers’ living conditions but also allow slowing down ongoing SOC degradation.

In contrast to the LP and LC sub-indicators, which have a continuous basis with Landsat and Sentinel missions [65] and for which there are also further high-resolution maps available [66], the SOC sub-indicator still lacks good spatial and temporal coverage. Further, there are currently no sufficient datasets available to provide information about the management or the input for the SOC indicator. Thus, the SOC change is only approximated by the LC change sub-indicator, leading to a misbalance towards the LULC change in the overall SDG 15.3.1 indicator. At the moment of the analysis, the high spatial resolution SOC data by Innovative Solutions for Decision Agriculture (ISDA) based on [67] were not available. Further work should thus address this limitation and incorporate per availability high-resolution SOC data in the analysis, as well as conducting field validation of both approaches. At the beginning of 2021, the UNCCD updated the first version of the SDG 15.3.1 good practice guidance and innovated the methodology [68]. Future studies should therefore adopt this new approach in conjunction with newly available datasets.

The improvement of the subnational analysis with freely available data, the use of cloud computing platforms, and the source code’s availability to perform LD assessment present an opportunity to upscale the analysis further and transfer the methods to other study areas.

5. Conclusions

The presented study demonstrates the potential of earth observation for LD monitoring with high spatial resolution data and uses cloud computing approaches with Google Earth Engine, and it improves the measurement of the SDG 15.3.1 indicator in the study area in Tanzania up until 2015 and 2019 at two different levels of spatial detail. Our study thus offers advancement of sub-national assessments of land degradation (LD) in heterogeneous landscapes. The improvement of the sub-national analysis with high-resolution data, the use of cloud computing platforms and the provision of the source code used here to perform LD assessment should encourage a transfer of the here presented approach to other study areas and/or the upscaling of the results of this study to the national level.

For this, we compared two approaches of assessing the SDG indicator 15.3.1 in Kiteto and Kongwa districts of Tanzania. The first method applied the global default (DM) medium resolution datasets proposed by the UNCCD for monitoring LD for the baseline period (BP, 2000–2015). The second method, the adapted method (AM), applied local land cover 30 m maps and 30 m Landsat to monitor LD for the baseline and the first monitoring period (MP, 2015–2019). The LD assessment for the BP reveals large differences between
the DM and AM. Using the DM, nearly all degraded area stems from the LP sub-indicator based on 250 m MODIS imagery. In contrast, the degradation was less than 1% for the LC and SOC change sub-indicators, calculated based on ESA CCI LC (300 m) maps. The LD captured by the AM based on Landsat time series and 30 m LC data was evenly distributed between the three sub-indicators and revealed LD on 27.7% of the area. We, therefore, concluded that the results derived from medium-resolution datasets are likely to over- and underestimate the LD for different sub-indicators and, thus, might misinform policy- and decision-makers and land managers if used operationally. Further, our study concluded that the local datasets and high-resolution imagery are essential to capture subtle changes within the heterogeneous landscape in semiarid central Tanzania.

Our results confirmed that LD is currently ongoing in the study area. The LD did not halt after 2015 but spread further across the districts and formed several severe LD clusters. Therefore, to achieve the national LDN targets, it is crucial to address the most important LD causes, such as overgrazing and unsustainable farming in the study area. The application of SLM practices would enhance the low LP in croplands and prevent LULC change in KK districts.

Further work should incorporate high-resolution SOC data in the analysis and conduct field validation of LD assessments resulting from both approaches.

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Appendix A

Table A1. Land cover transition matrix (2000–2015) based on the default methods (DM). Green, beige and brown colors indicate improving, stable and declining conditions of land cover categories, respectively. The area in km$^2$ and the possible cause of the land cover transition are indicated in the matrix. The change is based on the moderate-resolution land cover dataset.

| DM Land Cover Category in 2015 (km$^2$) | Forestland | Grassland | Cropland | Wetland | Urban | Otherland | 2000 Total (km$^2$) |
|---------------------------------------|------------|-----------|----------|---------|-------|-----------|-------------------|
| Forestland                            | Stable     |           |          |         |       |           | 2815.19           |
| Grassland                             | Afforestation 127.43 | Stable 6976.03 | Agricultural expansion 18.72 | Inundation 0.74 | Deforestation 0.74 | Vegetation loss 0 | 7122.92           |
| Cropland                              | Afforestation 3.09 | Withdrawal of agriculture 4.2 | Stable 6606.86 | Inundation 0 | Deforestation 0.49 | Vegetation loss 0 | 6614.40           |
| Wetland                               | Woody encroachment 0 | Waterbody drainage 0 | Waterbody drainage 0 | Stable 537.2 | Waterbody drainage 0.12 | Vegetation loss 0 | 537.34           |
| Urban                                 | Afforestation 0 | Vegetation establishment 0 | Agricultural expansion 0 | Wetland establishment 0 | Vegetation loss 1.54 | Withdrawal of settlements 0 | 1.54           |
| Otherland                             | Afforestation 0 | Vegetation establishment 0 | Agricultural expansion 0 | Wetland establishment 0 | Urban expansion 23 | Stable 0 | 0           |
| 2015 total (km$^2$)                   | 2940.71 | 6984.06 | 6626.25 | 537.96 | 2.41 | 0 | 17,091.40 |

Table A2. Land cover transition matrix in km$^2$ (2015–2018) based on the adapted methods (AM). Green, beige and brown colors indicate improvement, stable and decline of land cover category, respectively. The area and the possible cause of the land cover transition are indicated in the matrix. The change is based on the high-resolution land cover dataset.

| AM Land Cover Category in 2018 (km$^2$) | Forestland | Grassland | Cropland | Wetland | Urban | Otherland | 2015 Total (km$^2$) |
|---------------------------------------|------------|-----------|----------|---------|-------|-----------|-------------------|
| Forestland                            | Stable 1819.5 | Vegetation loss 89.5 | Deforestation 94.2 | Inundation 19.3 | Deforestation 5.4 | Vegetation loss 13.1 | 2041           |
| Grassland                             | Afforestation 26.2 | Stable 6993 | Agricultural expansion 368.1 | Inundation 77 | Urban expansion 22.7 | Vegetation loss 77 | 7563.9           |
| Cropland                              | Afforestation 9.3 | Withdrawal of agriculture 175.4 | Stable 4602.4 | Inundation 50.9 | Urban expansion 23 | Vegetation loss 64.6 | 4925.6           |
| Wetland                               | Woody encroachment 0.7 | Waterbody drainage 5.5 | Waterbody drainage 7.4 | Stable 173.7 | Waterbody drainage 0.8 | Vegetation loss 4.7 | 192.9           |
| Urban                                 | Afforestation 0.1 | Vegetation establishment 2.2 | Agricultural expansion 10 | Wetland establishment 0.6 | Wetland establishment 0.6 | Vegetation loss 1.4 | 211           |
| Otherland                             | Afforestation 2.7 | Vegetation establishment 55.7 | Agricultural expansion 69.7 | Wetland establishment 30.6 | Urban expansion 10.1 | Stable 1988.3 | 2157           |
| 2018 total (km$^2$)                   | 1858.4 | 7321.4 | 5151.7 | 352.2 | 258.7 | 2149.1 | 17,091.4 |
Figure A1. The land productivity component trend generated using (A) the default approach with MODIS imagery, (B) the adapted approach with Landsat imagery for the baseline period, and (C) the adopted approach with Landsat imagery for the monitoring period 2015–2019.

Figure A2. The land productivity component state generated using (A) the default approach with MODIS imagery, (B) the adapted approach with Landsat imagery for the baseline period, and (C) the adopted approach with Landsat imagery for the monitoring period 2015–2019.
Figure A3. The land productivity component performance generated using (A) the default approach with MODIS imagery, (B) the adapted approach with Landsat imagery for the baseline period, and (C) the adopted approach with Landsat imagery for the monitoring period 2015–2019.

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