Microscale and Mesoscale Aeolian Processes of Sandy Coastal Foredunes from Background to Extreme Conditions

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Abstract: Aeolian transport affects beach and foredune pre-storm morphologies, which directly contribute to storm responses. However, significant spatiotemporal variation exists within beach-dune systems regarding how biotic and abiotic factors affect topography. There are multiple metrics for quantifying topographic change, with varying pros and cons, but understanding how a system changes across spatiotemporal scales relative to varying forcings is necessary to accurately model and more effectively manage these systems. Beach and foredune micro- and mesoscale elevation changes (Δz) were quantified remotely and in situ across a mid-Atlantic coastal system. The microscale field collections consisted of 27 repeat measurements of 73 elevation pins located in vegetated, transitional, and unvegetated foredune microhabitats over three years (2015 to 2018) during seasonal, event-based, and background wind-condition collections. Unoccupied aerial System (UAS) surveys were collected to link microscale point Δz to mesoscale topographic change. Microscale measurements highlight how Δz varies more pre- to post-event than seasonally or monthly, but regardless of collection type (i.e., seasonal, monthly, or event-based), there was lower Δz in the vegetated areas than in the associated unvegetated and partially vegetated microhabitats. Despite lower Δz values per pin measurement, over the study duration, vegetated pins had a net elevation increase of ≈20 cm, whereas transitional and unvegetated microhabitats had much lower change, near-zero net gain. These results support vegetated microhabitats being more stable and having better sediment retention than unvegetated and transitional areas. Comparatively, mesoscale UAS surfaces typically overestimated Δz, such that variation stemming from vegetation across microhabitats was obscured. However, these data highlight larger mesoscale habitat impacts that cannot be determined from point measurements regarding volumetric change and feature mapping. Changes in features, such as beach access paths, that are associated with increased dynamism are quantifiable using mesoscale remote sensing methods rather than microscale methods. Regardless of the metric, maintaining baseline data is critical for assessing what is captured and missed across spatiotemporal scales and is necessary for understanding the contributors to heterogeneous topographic change in sandy coastal foredunes.

Keywords: beach access path; coastal management; erosion and accretion; plant presence/absence; structure from motion (SfM); tropical and extra-tropical cyclones; UAS; vegetation stabilization; wind climatology

1. Introduction

Coastal foredunes are inherently dynamic interface habitats that are increasingly vulnerable to change. Rising sea levels and the increasing frequency, severity, and unpredict-
ability of storms associated with climate change directly threaten these habitats and indirectly threaten those they protect as land–sea buffers [1,2]. Foredunes, defined as the first backshore, shore-parallel, vegetated dune ridge formed by aeolian sand deposition [3], are the first line of defense for inland features and infrastructure during high tides and storms. Despite their geological instability, these areas support increasing population sizes, infrastructure development, recreational activities, and economic investment [4–7]. Storm surge flooding and dune collision erosion [8] remain the main focus of assessing storm-induced changes [9–14]. However, wind events can significantly alter dune topography [14–17] and are predicted to contribute to increased erosion events in the future [18].

Aeolian transport affects dune habitat storm impact, which is largely dependent on the height of a foredune relative to storm surge [19–21]. Pre-storm dune elevation, volume, and shape are sensitive to both aeolian input and wave dissipation rates [22]. Storms may cause dune collision erosion and overwash, but between storm events, wind transport ultimately shapes topography to its next pre-storm state [8,14,16,17,19]. More powerful wind events garner the most research attention [13,23], in part because they tend to be associated with collision erosion and inundation, although they may also co-occur with rain, thereby limiting transport [24–26]. Significant aeolian transport from the beach to dunes is associated with low- to medium-magnitude wind events [26], with both wind speed and direction ultimately impacting transport [17,27,28].

Topography and obstructions (e.g., plants, structures) steer near-surface winds, altering transport. Wind velocity typically slows at the toe, accelerates upslope, then decreases at the crest and leeward slope [29–31]. Within the dune, transport decreases inland [32], and features like paths can funnel winds creating conditions for accelerated erosion [11]. Aboveground biomass can capture sediment and increase deposition even at 1% cover [33–37], affecting both wind and wave transport [23,38]. Foredune volume and vegetation cover are positively related [39], but more vegetation does not necessarily trap more sand [35,37,40]. In blowout and washover fans lacking vegetation [41–44], point measurements of aeolian transport and elevation change have been made [29,42–45]. However, across the broader beach–dune system, reported changes are more typically volumetric and measured via remote sensing [13,14,26], but variable sediment erosion and accretion across a system drive topographic heterogeneity [24,25].

Wind direction and intensity can vary seasonally, altering sediment transport between and among events. For example, in the Mid-Atlantic US, northeasterly winds predominate September to October, with hurricane season being from 1 June to 30 November [46]. Wind velocity also tends to be most seasonally mild in the Pacific Northwest US in the summer months [47]. As a result, milder elevation changes may be expected over summers such as in a Mediterranean and US dune system where elevation pins across a growing season showed <6 cm and <10 cm net accumulation April to September, respectively [48,49]. Annual net erosion or deposition may also be mild (less than 10 cm) as exemplified by the lacustrine dunes of Michigan, USA, from 1990 to 1998, where foredunes experienced a net deposition of 1–6 cm [50]. However, dune elevation is rarely stable, as one event can alter topography drastically, inducing 58 to 90 cm elevation gains in a year [12,13,38,51], with drastic localized changes in microhabitats [49]. Varying sediment transport ultimately impacts plant emergence and zonation as additional drivers of topographic heterogeneity [52].

Several approaches exist to remotely monitor the coastal beach system such as satellite orthoimagery, airborne lidar, unoccupied aerial vehicles (UAS), terrestrial laser scanners, stationary cameras, radar, ground-penetrating radar, and more [14,53–55]. Each has positive and negative aspects regarding the accuracy and ability to also quantify or compare to ecological data. When to use a particular method or combination of methods typically depends upon the size of the site, frequency of observations required, and spatial resolution of the resulting data [14]. Additional factors for choosing a remote sensing method include cost of the platform, sensors, processing software, and training. Perhaps
an underrepresented factor includes the consideration and inclusion of stakeholders and end-users of these datatypes, who need to be considering when choosing a platform, sensors, and data type to encourage the use of the information as well as sustainability of the method or dataset by management, conservation, and monitoring groups.

Understanding how dune systems change across different spatiotemporal scales relative to biotic and abiotic aeolian forcings is necessary to more accurately model future scenarios and better manage these systems [5,56,57]. Towards this goal, beach and foredune elevation changes were quantified across a Mid-Atlantic coastal system from 2015 to 2018 as seasonal, event-based, and background wind conditions. Remote sensing UAS surveys and more frequent field point measurements within vegetated, transition, and unvegetated barren foredune microhabitats were made. Mesoscale habitat level change is assessed relative to targeted change at localized points with associated physical and biological attributes ascribed. The point measurements are straightforward to collect and maintain, as remote sensing measurements are more computationally expensive [14], but both change metrics are options for coastal practitioners in assessing landscape states.

2. Materials and Methods
2.1. Study Site

Island Beach State Park (IBSP), is a high barrier island [58] located along the Barnegat Bay Peninsula, Berkeley Township, and Ocean County, NJ, USA (Figure 1). It is a microtidal environment containing ≈17 km of sandy beach shoreline transitioning the full range of barrier island habitats from littoral to dune, maritime forest, and tidal marsh. The mean grain size is medium quartz sands, 0.300–0.350 mm. The foredunes are dominated by native *Ammophila breviligulata* or invasive *Carex kobomugi*, with minor exceptions where *Toxicodendron radicans* is prevalent [38]. The growing season runs from May to October and annual precipitation is ≈127 cm. Precipitation and wind speeds are seasonally lowest April through August, with prevailing winds blowing from the southwest. Conversely, northerly winds and storms in the fall and winter have the greatest transport potentials [17,29]. Hurricane Sandy (October 2012) is the most recent major storm to impact IBSP [10]. The park recreation area is bounded by a north and south natural area (≈1.5 km) that are largely unmanaged. Between these areas is the recreational area of the park where prior to Sandy, the foredune was largely continuous, excluding the 24 beach access paths through the system. Here, Hurricane Sandy erosion was variable, increasing south to north latitudinally [38] with beach access 1 (A1) being the most northerly.

Field collections were performed among the foredunes A15–A24 (≈3 km). Here, collision erosion from Hurricane Sandy occurred, but the foredunes were not overwash or inundated [8]. The foredunes north of A15 experienced overwash and inundation and were largely destroyed. High wind speeds from Hurricane Sandy created blowouts and exacerbated the size and depth of previously existing blowouts in the foredunes and extending into the secondary dunes [38,49]. Over five years, as much as 2 m of deposition was documented in blowouts at IBSP with variation within and among these features [29]. Since Hurricane Sandy, the system had been recovering with minimal disturbances until fall 2016, when Hurricane Joaquin caused collision erosion and high winds [13].
Figure 1. Island Beach State Park, NJ collection area. Data collection consisted of quantifying changing elevation at point locations in three foredune microhabitats: in blowouts ( unvegetated), at 75 m and 150 m spacing along the foredune in its center (vegetated), and at the periphery of blowout edges, half vegetated and half barren (transitional). The background satellite imagery is USDA NAIP 1 m orthoimage from June 2015.

2.2. Microscale Metrics: Elevation Pins

A total of 73 elevation pins (i.e., erosion pins [29,43,44]) were installed in three foredune microhabitats and their heights, relative to the surface, were monitored from 2015 to 2018 (Figure 2). The pin microhabitats vary within the foredune ( blowout center, blowout periphery, or foredune center) and vegetation cover ( unvegetated, transitional, or vegetated). Transitional pins being at blowout peripheries were half vegetated, one side was barren, and the other vegetated. All pins were sturdy (fiberglass or steel), driven ≥ 15 cm into the sand, and GPS-marked. Data consists of repeat measures of pin height over time (sediment surface to pin top), to examine elevation change (Δz) between successive measurement times referent to as collections. Negative and positive Δz values indicate erosion and accretion, respectively and Δz could only be determined for pins present in both collection Tx and Tx+1; the sample sizes presented in Table 1 reflect instances of Δz not being possible to compute between collections. During collections, if a pin was missing, then its location was dug to ensure it was not buried; if buried, then burial depth relative to surface height was recorded and a new pin was installed. A new pin was also installed for any missing pins, and for any pins within 0.3 m of being buried.
An initial 20 pins were installed on 10 June 2015, in the blowout centers (unvegetated, installed in the flat center). These were 1.37-cm diameter galvanized steel, placed at 1 m height above the surface. Initially, collections were event-driven, as a proof-of-concept, then were monthly for 16 months (February 2016 to June 2017, collection 3 to 19), then seasonally (beginning and end of a season) and event-based until the final collection in the fall of 2018 (Table 1 and Figures 1 and 2). Event-based collection dates were opportunistic and based upon weather predictions pre-storm, reports of measured wind velocities post-storm, or post-event wind and wave data from NOAA Tides and Currents, the National Center for Environmental Prediction, and the National Hurricane Center (NHC). Over the study duration, there were 27 total collections for the unvegetated pins.

Table 1. Dates of all 27 total collections for the elevation pin types installed throughout the IBSP foredunes. The number in each cell represents the sample size of each period reflecting pins going missing or becoming buried between collections. Blue cells denote the period number, Δz from one collection to the next, where period one is Δz between collections one and two.

![Figure 2](https://example.com/figure2.png)
m from the closest beach access path and the vegetation stands they were installed in were dominated by either invasive C. kobomugi (n = 14) or native A. breviligulata (n = 17). On collection 11, a final 22 pins were installed at the peripheral edge (east wall n = 20) of blowouts, representing the transition between the vegetated foredune and the unvegetated blowout. These were the only pins installed on a slope and the 22 stands were all dominated by A. breviligulata largely as pure monocultures except for one stand dominated by C. kobomugi and the occasional presence of a handful of secondary stabilizing species such as Seaside Goldenrod (Solidago sempervirens) or Sea Rocket (Cakile endulata; [49]). The full range of species found at each transect over time can be found in Charbonneau, 2019 [49]. These 53 total pins were 1.8-m fiberglass driveway markers driven 0.5 m into the ground. All replacement pins were of this making. Pictures of pins are available in Supplemental 1.

2.3. Mesoscale Remote Sensing Data Collection

Three UAS surveys were performed at 80 m altitude, and 80% and 75% along and across track overlap, respectively, with a DJI® Phantom 3 Advanced with standard camera payload or SenseFly eBee© RTK with SenseFly DSC-WX220 camera. Geotagged images were mosaicked using SfM software (Pix4D©, Agisoft© Photoscan, now Metashape) across the full collection area encompassing the foredune, beach, and swash zone [14]. UAS survey dates were 3 March 2017 (flight 1, DJI Phantom 3 Advanced) as plants broke dormancy ahead of a potential storm, December 11, 2017 (flight 2, SenseFly eBee RTK) after the end of the 2017 hurricane season, and April 22, 2018 (flight 3, SenseFly eBee RTK) following the winter storm season (Figure 2). Flights coincided with a pin survey. All flights were referenced with ground control points, every ≈100 m along the study extent, which were permanent road fixtures west of the foredune and temporary beach markers surveyed with a Topcon© GR-5 RTK GPS system [14]. Survey platform resolution and uncertainties are presented in Table 2. Uncertainties were calculated from in -objects (i.e., field vehicles, beach artworks, dumpsters, etc.). Publicly available lidar DEMs filled temporal gaps between UAS and microscale surveys (Table 2). All elevation datasets were transformed to New Jersey State plane coordinates before analyses and DEMs (1 m, gridded) were processed for elevation derivatives such as slope, profile curvature, and aspect using corresponding tools within ArcGIS Pro.

| Data | Source/Instrument | Dates (yyyy-mm-dd) | Resolution (Grid or Cell Size) | Uncertainty | Analysis |
|------|------------------|---------------------|--------------------------------|-------------|----------|
| satellite imagery | USDA NAIP | 7 June 2015 | 1 m | ±6.1 m | map background; mask creation |
| topobathy lidar DEM | USGS, NOAA, CMP, USACE | 2014, September 2017 | 1 m | ±0.09-0.15 m | mesoscale metrics |
| UAS DEM | DJI® Phantom 3 Advanced™, eBee™ SenseFly® | 21 March 2017, 11 December 2017, 22 April 2018 | 0.05 m | ±0.15 m | mesoscale metrics |
| mapped features | Trimble® GeoXT™ GeoExplorer® | 2013–2018 | 0.15 m | ±0.8 m | microscale metrics |
| sediment samples | coring, brab | 2018 | --- | --- | aeolian critical threshold |

2.4. Mesoscale Metrics Derived from Digital Elevation Models

Feature extraction of the dune’s crest location, mean higher high water (MHHW) line, beach width, crest elevation, and surface elevations, at all pins, occurred in ArcGIS Pro v2.8. The custom-created ArcGIS model extraction tools are detailed in Supplemental 2.
In the tool, crest line features were identified from the maximum positive profile curvature for the subaerial beach [59], and elevations were extracted along each survey every 0.2 m; this tool could be used to extract the dune toe modified for maximum negative curvature. NOAA tide gauge MHHW values (z = 0.61 m; station 8534720) were used to extract the MHHW location as polylines from subaerial DEMs. Beach width represents the distance between the crest and MHHW, as this area was consistently exposed to aeolian processes during non-precipitation periods. Masks of the subaerial study extents of beach and woody habitats were generated using visual satellite orthomagery delineation (Table 2). These masks aided extraction tools by decreasing false positives when searching for desired criteria and removing woody habitats. Each extraction method entailed minor buffering and smoothing conditions to remove erroneous values.

Elevation was extracted and Δz computed between flights at pin locations. DEMs were differenced to create a Δz surface where each grid was also categorically defined as having experienced erosion or accretion between flights. A spatial sensitivity analysis was performed to quantify the distance pin Δz agreed with categorical Δz (i.e., erosion or accretion) using the Distance and Direction tool to manually draw a radius from each pin on the categorical Δz surface to the nearest categorical switch.

Beach access-path impacts were quantified using a custom ArcGIS tool (Supplemental 3). The tool utilizes Δz surfaces between consecutive surveys to create standard deviation surfaces, where impact polygons are defined by isolating areas greater than one standard deviation from the mean (Figure S2.1). Beach path polylines (previously mapped [38]) were buffered by 12.2 m and used to extract areas from the standard deviation surface. The buffer size was based on local topography to exclude nearby blowouts. Polygons larger than 0.93 m², based on raster resolution, were analyzed using the Minimum Boundary Geometry tool to determine access the impact’s major and minor axes, surface area, orientation, and perimeter. These parameters are collectively referred to as access-impact geometries. Using the Δz surface, access impacts were categorized as erosional or accretional and polygons were manually identified as the head (inland), crest (seaward termination), or middle. Access impacts were quantified chronologically in three instances 2014 (lidar) to September 2017 (lidar), September 2017 to December 2017 (UAS flight two), and December 2017 to April 2018 (UAS flight three), and then across the full study period, 2014 to April 2018 (Table 2). March 2017 (UAS flight one) to September 2017 (lidar) values are not included in the analysis as the absolute error between these two surfaces was too large.

2.5. Sediment Composition and Wind Data

In the summer of 2018, 58 cores were collected to quantify grain distributions. A core was collected at each vegetated pin to a depth of 1 m. An additional 20 cores, sampled to 50 cm in depth, were collected from the backshore 10 m from the dune toe and longitudinally aligned with the 20 vegetated pins that were spaced 75 m apart in the southern survey half. All cores were collected with a T-core sampler of 30-cm tube length. Samples were dried, then sorted for 5 min in a RX-29 Ro-Tap® cascade shaker, using eight different US Standard Sieve Mesh Sizes, mesh #25 (0.71 mm, Φ0.5), #35 (0.50 mm, Φ1.0), #45 (0.35 mm, Φ1.5), #60 (0.25 mm, Φ2.0), #70 (0.21 mm, Φ2.25), #100 (0.149 mm, Φ2.75), #140 (0.105 mm, Φ3.25), and the bottom pan (<0.105 mm, Φ > 3.25). Raw sieve weights were processed using SANDY to generate mean grain size, kurtosis, and skewness values per sample [60]. See Supplemental 4 for more details on sediment distributions.

Wind velocity and weather (precipitation and temperature) data from 2015 to 2018 were collected from the nearest Rutgers NJ Weather Network station (Seaside Heights). The sensor is 14 km north of the northern study extent on the barrier island at 12 m elevation (~74.0738, 39.94383). These data are largely continuous, where data gaps (i.e., sensor outages) in one or more parameters were supplemented from the nearby Harvey Cedars or Sea Girt stations. See Supplemental 5 for more details on station sensors, outages, and
wind velocity presented as wind roses [61], produced in MATLAB® over the study duration [62]. Data are at 5-min intervals and corrected from sensor height to surface level using von Karman’s law of the wall equation [63].

\[ U_c = \left[ \frac{u_2}{\ln(z_0 / \kappa)} \right] \kappa \]  

(1)

\[ z_0 = 0.035 \times \ln \left( \frac{d}{0.18} \right) \]  

(2)

where \( z = \) sensor height in meters, \( U_c = \) wind velocity at the sensor height in m/s, \( z_0 = \) the grain size roughness length, \( \kappa = 0.4, d = \) sediment size in meters.

Wind data was categorically defined as coming from the northeast \((0° \text{ to } 90°)\), southeast \((91° \text{ to } 180°)\), southwest \((181° \text{ to } 270°)\), or northwest \((271° \text{ to } 360°)\) quadrant as well as onshore \((0° \text{ to } 180°)\) or offshore \((181° \text{ to } 360°)\). The number of instances whereby wind speed was at or above the critical velocity for entrainment was determined by first correcting for the elevation difference between the instrument and ground level [14]. Next, the critical Shield threshold for initiation of motion for the maximum \(d50\) (0.23 mm) at the site was calculated as:

\[ U_{ct} = A \sqrt{\frac{\rho_e - \rho}{\rho}} gd \]  

(3)

where \( A = 0.1, \rho = 1.225 \text{ kg/m}^3, g = 9.81 \text{ m/s}^2,\) and \((d = 0.2 \text{ mm, } \rho_i = 2650 \text{ kg/m}^3, \text{ for quartz sand}),\) not corrected for beach slope, air temperature, or humidity. Using Equation (3), the critical velocity was 0.22 m/s for the largest grains sampled (Bagnold, 1943). Wind roses were created using this critical velocity to determine when sediment motion was possible and delineated with respect to concomitant precipitation events, as well.

2.6. Statistical Analyses

Unless otherwise noted, all tests are two-tailed with \( \alpha = 0.05 \) and were performed with JMP Pro 15.0 or Microsoft Excel 2016. All means are ± standard error. All wind roses were created in MATLAB® [61,62]. \( \Delta z \) was typically evaluated as the absolute value (\(|\Delta z|\)), categorized as erosional or accretional, given that both erosive and accretive instances occurred concomitantly within and among the collections.

2.6.1. Microscale Analyses

Analyses on microscale \( \Delta z \) or \(|\Delta z|\) test its variability across seasons (fall, winter, spring, summer), collection types (event-based, seasonal, or monthly), and elevation pin types (vegetated, unvegetated, transitional). Pin types and collection durations used in the different tests balance sample sizes and time between collections relative to the hypotheses in question. Where noted, instances of no net change \((\Delta z = 0)\) were not included to prevent a loss of degrees of freedom. Instances of no net change were random and evenly distributed across the pin types. Comparisons between groups in multivariate analyses are Tukey HSD pairwise comparisons.

General Variation between Collection Types

ANOVA was used to examine if transition pin \(|\Delta z|\), between collections and over their install duration (collection 11 to 27) varied based on their position in the north or south side of the east blowout wall and by aspect (N, S, NE, or SE facing). Linear regression was performed to test if days between collections impact mean \( \Delta z \) period 1–26 for the unvegetated data. Fisher’s exact tests were used to test if categorical instances of erosion or accretion occurred more or less than would be expected by chance.
Month-to-Month Seasonal Variation Analyses

$|\Delta z|$ seasonal variability, month-to-month and in different seasons, was tested with full factorial multivariate regressions. The independent variables are change type (erosion or accretion) and season. This test was performed using only the monthly unvegetated data from collections 4 to 18—a full year—excluding instances of no net change, where periods 4–6 and 16–18, 7–9, 10–12, and 13–15 represent spring, summer, fall, and winter, respectively (Table 1); the test was rerun with period 10 excluded as a potential outlier.

Vegetation Impact Analyses

Seasonal variability was analyzed with pin type to test if and how vegetation impacted $\Delta z$. Monthly collections with all pin types present did not span a full year (Table 1), so that $|\Delta z|$ was analyzed from seasons’ start to end, as in collections 13–16, 16–19, 19–20, and 20–21, as the winter, spring, summer, and fall of 2017, respectively, and 22–25 and 25–26 as the spring and summer of 2018, respectively (Table 1). Three tests using combinations of the above data were performed to ensure results were robust across time: (1) all six seasons were used with uneven sample sizes, given more spring and summer replicates, (2) four 2017 seasons were used, and (3) collections from the winter and fall of 2017 were analyzed with those from spring and summer of 2018. The three tests were full-factorial multivariate regressions, excluding change type as an independent variable, given that it and interactions including it were not significant.

The vegetation impact was tested in event-based and monthly collections. A full-factorial multivariate regression was performed on $|\Delta z|$ with the independent variables of change type, pin type, and period (for periods 20, 23, 24, and 26, all three pin types were installed and observed an event); an event did not occur in period 22, so that period 22 and general instances of no change were excluded. The same full-factorial multivariate regression was also performed on $|\Delta z|$ in the monthly collections (period 11 to 18, all pin types present). Lastly, from the first collection with all pin types to the last such (collections 11 to 27), the difference in total $\Delta z$ was examined with ANOVA; outliers ($n = 7$), defined as two or more standard deviations from the mean for each pin type, were excluded. Examining the potential impact of different vegetation-cover types and densities is outside of the scope of this research but will be examined in a subsequent publication. For reference, the vegetated pins maintained a density of $107 \pm 9$ plants and $48\% \pm 2.8\%$ cover (1 m² transect, with the pin in the middle) where transition pins maintained similar density and percent cover in half the transect and were bare the other half.

2.6.2. Mesoscale Analyses

Mesoscale analyses link microscale and mesoscale metrics and quantify the impact of beach access paths on $\Delta z$. $\Delta z$ quantified from complimentary UAS (flights one to two, two to three, and one to three) and field data (collections 16 to 21, 21 to 24, and 16 to 24) were compared between the remote versus field data using Wilcoxon tests. Missing data from one dataset are excluded from the other to balance the sample sizes, accounting for instances of missing pins and areas of poor image stitching in the field and remote data, respectively. Outliers, defined as having $\Delta z$ greater than 150 cm relative to the vertical uncertainty of the remote datasets were excluded from these tests. Kruskal–Wallis tests were performed per flight period to compare spatial sensitivity radii across pin types.

Beach impacts were categorized as $\Delta z$ accretional or erosional at access impact areas and were compared with Mann–Whitney tests against the access impact geometries. Access impact locations (crest, head, middle) were examined as impacting geometries using Kruskal–Wallis tests. Beach-access impact geometries are defined as the major axis orientation relative to true north, polygonal minor axis distance (i.e., width), polygonal major axis distance (i.e., length), polygonal perimeter distance, and polygonal surface area.
3. Results

3.1. Site Characteristics during Survey Periods

Mean beach width and standard deviation decreased overall from 2014 to April 2018 (72 ± 255 m to 63 ± 150 m, respectively; Figure S2.43). This decrease corresponded with an overall shallowing of the mean foreshore slope from 40–50° to 0–15°, with the foreshore slope aspect, relative to true north, ranging from 67.5° to 112.5°. The study area experienced a mean accretional Δz of +0.56 ± 2.42 m and normalized volumetric change of +0.12 ± 0.003 m³/m² from 2014 to April 2018.

The site experienced wind speeds greater than critical sediment mobility for 3.8% of the 4-year records. The majority of mobility periods, 83%, occurred during times of no precipitation and occurred in two main directional sectors: 45–135° (NE-SE, stronger speeds) and 285–10° (NNW-NNE, more frequent). These sectors align with the predominant winter storm (extra-tropical cyclones) and tropical cyclone wind directions in this area. See Supplemental 5 for additional details of the aeolian analysis.

3.2. Microscale Measured Variation between Collection Types

When all three pin types were installed, period 11, on both Δz (F2,107 = 15.26, p < 0.0001) and |Δz| (F2,107 = 18.11, p < 0.0001) varied by collection type and were greater in event-based than monthly and seasonal collections. Event-based collections were primarily more accretional than would be expected by chance, whereas monthly collections were primarily erosional, and seasonal collections did not have more or fewer instances of erosion or accretion than would be expected by chance (Fisher’s Exact Test, p < 0.0001). Transition pin location on the east wall and aspect did not impact Δz (p > 0.05) and Δz between collections was unrelated to time between collections in unvegetated pins (R² = 0.10, p > 0.05).

3.3. Microscale Measured Month-to-Month Seasonal Variation

Unvegetated pins’ month-to-month |Δz| variation was seasonally dependent (F3,170 = 10.73, p < 0.0001). They also varied between erosional and accretional occurrences (F1,170 = 11.04, p < 0.0001), with a significant season X change-type interaction (F3,170 = 5.68, p = 0.0001) stemming from greater |Δz| in fall (12.5 ± 3.1 cm) than in other seasons (4.7 ± 0.5 cm), where the differences in accretional |Δz| (9.35 ± 1.17 cm) versus erosional instances (4.28 ± 0.98 cm) were driven by fall collections. These results remained true when excluding period 10 as a potential outlier.

Seasonal differences, in the above unvegetated analyses, stemmed from wind direction in instances above critical velocity (Figure 3). Mean wind speed during these periods was similar from all directions (NE, SE, NW, and SW) with speeds between 0.25–0.28 m/s and the greatest wind speeds having occurred in spring and winter. Fall was the only time when northeast winds were predominant (44% of instances) and of greater magnitude, compared with other seasons, when northwestern winds were predominant (40–54% of instances; Figure 3A). Analyzing fall wind speed in (period 10–12), mean Δz in period 10 (20.03 ± 4.69 cm), when northeastern winds predominated, was more than double period 11 (7.58 ± 4.06) and 12 (4.94 ± 3.83), despite equivalent wind speeds from other directions (Figure 3B). Spring 2016 and 2017 had nearly identical wind directions, but greater wind speeds in 2017, yet equivalent Δz (R² = 0.31, p = 0.75; Figure 3A,C). Northeastern winds encompassed 22%, 30%, 44%, and 23% of instances in spring, summer, fall, and winter, respectively.
3.4. Impacts of Vegetation on Microscale Measurements

Variation in $|\Delta z|$ across seasons, by pin type, consistently supports vegetation as reducing sediment movement (i.e., increasing stability; Table 3). In all three seasonal tests using different temporal data, $|\Delta z|$ varied by pin type, with transition pins not varying from either the vegetated or unvegetated pins, but the vegetated pins having lower $|\Delta z|$ than unvegetated pins (Figure 4A; Table 3). Pairwise differences between the seasons varied with the data used in tests 1, 2, and 3.

Figure 4. Variation in $|\Delta z|$ between pin types. Across the three types of collections, (A) seasonal, (B) event-based, and monthly (C) there was consistently less $|\Delta z|$ in the vegetated pins compared with the unvegetated and transitional pins. Mean differences across pin types hold, regardless of whether $|\Delta z|$ was equivalent between instances of erosion versus accretion (e.g., seasonal and monthly) or accretion was of a greater magnitude (e.g., events). Error bars are standard error and pin types with different letters denoting a significant difference between pin types (Tukey HSD).
Table 3. Variation in |Δz| across seasons by pin type consistently supports vegetation as reducing sediment movement with pairwise differences across seasons, based on which subset of the temporal data was analyzed. Analyses were full factorial multivariate regressions. Season and pin type never had a significant interaction (p > 0.05).

| Variation in | Statistical Test | Pairwise Comparisons |
|-------------|------------------|----------------------|
| Seasonal Test 1: All 6 seasons | | |
| transitional = unvegetated and vegetated pins; | F\(_{2,395} = 8.35, p < 0.001\) | fall > spring and summer |
| | F\(_{2,63} = 4.91, p < 0.01\) | No pairwise seasonal differences |
| Seasonal Test 2: 2017 data | | |
| vegetated < unvegetated | F\(_{2,262} = 3.89, p = 0.02\) | spring > summer and Fall |
| Seasonal Test 3: 2017 and 2018 Combined | | |
| vegetated < unvegetated | F\(_{2,262} = 12.70, p < 0.0001\) |

Variation in |Δz| during event-based and monthly collections supported vegetation stabilization. During events, |Δz| varied by pin type (F\(_{2,218} = 10.89, p < 0.0001\)) with vegetated pins having less |Δz| than both the transition and unvegetated pins (Figure 4B) and the magnitude of |Δz| was greater in instances of accretion than erosion (F\(_{1,218} = 4.56, p = 0.03\)). |Δz| did not vary between the four event periods (p = 0.83). In monthly collections, vegetated pins had less |Δz| than both the transition and unvegetated pins (F\(_{2,499} = 12.81, p < 0.0001\); Figure 4C) and the magnitude of |Δz| was equal in instances of erosion and accretion (p = 0.29). There were no significant interactions in either the seasonal or monthly multivariate regression tests.

Across the study duration, from the first collection with all three pin types to the final collection (11 to 27), Δz varied by pin type (F\(_{2,29} = 8.51, p < 0.001\), with vegetated pins having greater Δz than both the transition and unvegetated pins (Figure 5). Variation in Δz across collection and pin types are presented in Table 4. Instances of erosion and accretion were equally spread across the pin types, as would be expected by chance in monthly and seasonal collections, but the monthly collection transitional and unvegetated pins had more instances of erosion and fewer instances of accretion than would be expected by chance, whereas the opposite was true for vegetated pins in these same monthly collections (Fisher’s exact test, p < 0.0001).

Figure 5. Total Δz, from the first collection with all three pin types to the final collection (collections 11 through 27, from October 2016 to September 2018). Despite less incremental Δz, erosional or accretional, season to season and during events throughout their installment, the vegetated pins accreted more than both the unvegetated and transitional pins. Error bars are standard error.
Table 4. Vegetated pins consistently had less Δz (cm), regardless of collection type, but greater Δz increase over the duration of the install (Total Δz) of all three pin types, collection 11 to 27. Differences between Δz and |Δz| highlight the importance of performing analyses on |Δz| given that instances of both positive and negative values in Δz otherwise obscure the results. Units are cm and means are presented with standard error.

| Pin Type | Collection Type | Min Δz | Max Δz | Mean Δz | Mean | Accretion Mean Δz | Erosion Mean Δz |
|----------|-----------------|--------|--------|---------|------|------------------|----------------|
| u        |                 | -47.4  | 106.7  | 11.2 ± 2.6 | 17.2 ± 2.3 | 23.0 ± 3.5        | -10.3 ± 1.8     |
| transitional |      | -37.2  | 68.6   | 4.2 ± 1.3  | 7.5 ± 1.1   | 10.4 ± 1.7        | -5.9 ± 1.5      |
| vegetated |                 | -16.5  | 40.6   | 4.3 ± 0.7  | 5.4 ± 0.6   | 7.0 ± 0.8         | -4.6 ± 1.2      |
| unvegetated |   event  | -30.1  | 7.7    | -1.8 ± 0.9 | 3.6 ± 0.7   | 2.7 ± 0.6         | -5.5 ± 1.3      |
| transitional | seasonal | -40.6  | 29.6   | 1.6 ± 1.0  | 5.0 ± 1.1   | 3.9 ± 1.1         | -6.9 ± 2.0      |
| vegetated |                 | -54.6  | 33.7   | 2.0 ± 0.8  | 4.4 ± 0.8   | 5.8 ± 1.0         | -6.1 ± 2.4      |
| unvegetated |          | -37.3  | 71.1   | -0.5 ± 0.6 | 4.6 ± 0.5   | 7.1 ± 1.4         | -5.0 ± 0.6      |
| transitional | monthly | -36.8  | 35.6   | -1.0 ± 0.7 | 5.3 ± 0.5   | 5.4 ± 0.9         | -6.2 ± 0.7      |
| vegetated |                 | -17.8  | 27.9   | 0.6 ± 0.3  | 2.4 ± 0.2   | 3.1 ± 0.3         | -3.1 ± 0.4      |
| unvegetated |          | -40.0  | 41.9   | -4.8 ± 8.3 | 28.2 ± 3.1  | 27.3 ± 4.1        | -28.9 ± 4.6     |
| transitional | total Δz  | -25.4  | 35.6   | -0.8 ± 16.9 | 13.4 ± 2.4  | 11.1 ± 3.7        | -16.2 ± 2.8     |
| vegetated |                 | -45.7  | 61.6   | 22.0 ± 4.1 | 25.8 ± 3.2  | 26.6 ± 3.3        | -19.1 ± 13.5    |

3.5. Microscale to Mesoscale Metric Linking

The mesoscale UAS data largely differed from field pin collections in accurately quantifying the direction and magnitude of Δz (Figure 6). Categorically, the remote sensing data were in agreement with the field measurements for the type of change occurring (i.e., erosion, accretion, or no change) in 71%, 51%, and 71% of instances over flights one to two, two to three, and one to three, respectively. Correct versus incorrect instances were evenly split across pin types, varying slightly by survey period. In instances of categorical congruence and incongruence between the two data types, the UAS surface overestimated Δz flight one to two and one to three. In flights two to three, the UAS surface underestimated Δz compared reality in instances of categorical incongruence and was equal for instances of congruence (Table 5).

Table 5. Variation in Δz calculated from the UAS surface versus pins between flights.

| Flights 1 to 2 | UAS Surface Δz vs. Actual (Pin) Δz | Mean UAS Δz | Mean Pin Δz | Wilcoxon Test Result |
|----------------|-----------------------------------|-------------|-------------|---------------------|
| instances of categorical incongruence between data types | UAS surface overestimated Δz | 28.96 ± 23.3 cm | -7.83 ± 6.11 cm | Z = 2.31, p = 0.02 |
| | UAS surface underestimated Δz | 23.85 ± 21.88 cm0.70 ± 6.05 cm | Z = 0.84, p = 0.4 |
| Flights 1 to 3 | instances of categorical incongruence between data types | -8.39 ± 4.14 cm | 8.23 ± 2.25 cm | Z = -4.28, p < 0.0001 |
| Flights 1 to 2 | instances of categorical congruence between data types | UAS surface overestimated Δz | 56.88 ± 13.13 cm4.51 ± 3.27 cm | Z = 4.26, p < 0.0001 |
| Flights 1 to 3 | UAS surface underestimated Δz | 46.78 ± 12.48 cm13.41 ± 4.23 cm | Z = 4.15, p < 0.0001 |
| Flights 2 to 3 | instances of categorical congruence between data types | UAS surface & actual equal | 2.77 ± 2.83 cm | Z = 4.15, p < 0.0001 |
Figure 6. The UAS surfaces typically overestimated the magnitude of Δz (cm) relative to the field elevation pin measurements, regardless of whether the two data types were in agreement with the type of change-erosional or accretional–that occurred. Over-estimation ranged from 3–34x the true magnitude observed in situ. Slight differences in collection times could theoretically account for some of these differences, but not to the degree observed given that microscale data show minimal changes in elevation in the absence of an event and there were no events between the temporal incongruence of the datasets.

Differences in Δz across pin types (i.e., attributable to vegetation impacts) were not observable in the mesoscale data. Similarly, looking at microscale Δz, as a whole, from one flight to the next showed equivalent Δz across pin types, stemming from the high standard error associated with the means per group (p > 0.05).

Across flights, the spatial sensitivity analyses largely showed agreement between pin types regarding the distance beyond a pin at that Δz switched categorically from erosional to accretional or vice versa. The datasets maintain the same change category for $\bar{x} = 16.48 \pm 1.15$ m and $\bar{x} = 13.73 \pm 1.09$ m over flights one to two and one to three, respectively, with no differences across pin types (p > 0.05). Only for flights two to three does the range of data sensitivity vary across pin types (Figure 7).

Figure 7. A zoomed view (accesses 19 to 17) of spatial sensitivity analysis showing measured buffers (yellow circles) around numbered elevation pins (symbols) overlain on a categorical Δz map from flights two to three, April 2018 to December 2017. Forested areas were masked and excluded. Only between flights two to three was there a difference in radii, with vegetated pins showing a wider

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radius than both the transitional and unvegetated areas (Kruskal–Wallis Test, \( x_\bar{V} = 4.57 \pm 0.78; x_\bar{R} = 1.75 \pm 0.47; x_\bar{U} = 1.37 \pm 0.52; X^2 = 13.64, p = 0.001 \). (Kruskal–Wallis Test, \( x_\bar{V} = 4.57 \pm 0.78; x_\bar{R} = 1.75 \pm 0.47; x_\bar{U} = 1.37 \pm 0.52; X^2 = 13.64, p = 0.001 \)).

3.6. Beach Access Path Impacts on Local Topographic Change

Throughout the study, from 2014 to April 2018, 38 individual polygons were identified as increased areas of dynamism linked to accesses (37% erosional and 63% accretional). Of the erosional areas, 15.8%, 13.2%, and 7.9% occurred at the path’s crest, middle, and termination head, respectively. Of the accretional areas, 18.4%, 28.9%, and 15.8% occurred at the path’s crest, middle, and termination head, respectively. The identified access impacts began at the crests in 2014 and migrated inland towards the heads of each access path by April 2018.

Access-impact geometries shifted over each survey period. The mean width and length of impact areas increased on each subsequent survey, while orientation shifted eastward (Table 6). Sinuosity calculations for the 12 access paths generated a mean value of 1.05, indicating linear paths. The number of impact areas decreased from the mesoscale survey periods one to two, while their sizes (width, length, and surface area) increased (Table 6), suggesting the impact areas coalesced into fewer, larger areas. In the subsequent mesoscale period (period three), it is not clear whether, relative to period two, the previously existing impact areas separated and grew or additional impact areas formed (Table 5).

Table 6. Mean ± standard error for access impact polygon geometries during the survey periods, 2014 to April 2018. Orientation is reported in degrees from true north, and surface area as the summed surface area of the identified impact polygons.

| Survey Period | 1: 2014 to September 2017 | 2: September 2016 to December 2017 | 3: December 2017 to April 2018 | 4: 2014 to April 2018 |
|---------------|---------------------------|----------------------------------|-------------------------------|----------------------|
| impact polygons identified (n) | 34 | 24 | 35 | 38 |
| width (m) | 5.4 ± 0.9 | 8.1 ± 1.3 | 8.9 ± 1.2 | 9.9 ± 1.3 |
| length (m) | 11.8 ± 1.8 | 18.3 ± 3.4 | 17.4 ± 2.5 | 27.6 ± 2.4 |
| surface area (m²) | 2677 | 4515 | 6084 | 11,610 |
| orientation | 77.2 ± 8.8 | 84.6 ± 9.2 | 86.4 ± 7.8 | 90.7 ± 7.9 |

Paths’ crest, middle, and head areas had largely equal impact-geometry metrics. In mesoscale period one (2014 to September 2017), impact areas occurred at all paths’ crests and middles, while no effects were identified at paths’ inland termination heads (Figure 8); specifically, crests had larger impact areas—more than double, in most instances—than paths’ middles for all geometries, except orientation, which did not vary locally along paths: width (\( \bar{x}_C = 25.3 \pm 5.0 \text{ m}, \bar{x}_M = 11.6 \pm 2.7 \text{ m}, \bar{Z} = 2.19, p = 0.029 \)), length (\( \bar{x}_C = 56.5 \pm 10.2 \text{ m}, \bar{x}_M = 24.4 \pm 5.1 \text{ m}, \bar{Z} = 2.47, p = 0.014 \)), perimeter (\( \bar{x}_C = 137.2 \pm 23.4 \text{ m}, \bar{x}_M = 60.8 \pm 12.9 \text{ m}, \bar{Z} = 2.47, p = 0.014 \), and surface area (\( \bar{x}_C = 1428 \pm 373 \text{ m²}, \bar{x}_M = 360 \pm 175 \text{ m²}, \bar{Z} = 2.45, p = 0.014 \). For the other mesoscale survey periods, there were no difference between access impact geometries, locally, among the heads, middles, or crests of beach access paths.
Figure 8. Beach-access path impact geometries from 2014 to April 2018. Outlines indicate identified areas of increased elevation change greater than 1 standard deviation of the average elevation change for the difference surface. Red and blue shading indicates categorical elevation change type for whole study period of April 2018–2014. The left map (orange outline) is the northern most section, the center map (black outline) is the central access paths, and the right map (yellow outline) is the most southern access paths.

Paths’ impact areas varied in their geometries based on whether erosion or accretion had occurred, but not consistently across time. Only in mesoscale survey period three (December 2017 to April 2018) was there a difference in impact geometries between erosional and accretional polygons; specifically, impact area geometries were larger for instances of accretion than erosion in all metrics except orientation: length ($\bar{x}_A = 68.7 \pm 10.3$ m, $\bar{x}_E = 39.6 \pm 13.8$ m, $Z = 2.47, p = 0.014$), perimeter ($\bar{x}_A = 171 \pm 24.3$ m, $\bar{x}_E = 98.6 \pm 31.6$ m, $Z = 2.45, p = 0.014$), and surface area ($\bar{x}_A = 2342 \pm 597$ m$^2$, $\bar{x}_E = 1156 \pm 801$ m$^2$, $Z = 2.25$, $p = 0.024$). Elevation change category did not impact path geometries mesoscale survey periods, one, two, or four.

4. Discussion

The greatest elevation changes were associated with unvegetated blowout microhabitats in the event-based macroscale collections. Supporting this, $\Delta z$ was unrelated to time between collections and the greatest single pin $\Delta z$, 106.7 cm (Table 4), was in an event-based collection in a blowout, specifically associated with Hurricane Joaquin and a nor’easter event (mid-latitude extratropical cyclone) in fall 2015 [13,14]. The unvegetated $\Delta z$ values observed fit within the range of previously reported values for this microhabitat [12,13,38,51]. However, it is important to note that individual blowouts support heterogeneous abiotic conditions translating to heterogeneous transport potentials within them [42,43,46,64,65]. As a result, the installation of multiple pins within a blowout would be
expected to yield different $\Delta z$’s, as exemplified from the transitional pins typically having lower $\Delta z$ than the unvegetated pins. The unvegetated pin results match those of Harris who observed, over a year, elevation changes in three blowouts ranging from 1–107 cm [45]. In the three-year install, the mean net change in unvegetated pin elevation (6.2 ± 10.4 cm) compared to its range each collection (Table 4) suggests that elevation changes are likely ephemeral until or unless vegetation comes to recolonize and stabilize these microhabitats. This may be why blowouts, which are considered ephemeral features themselves, present at the study site in the 1990s may still be present today, 30+ years later [29,46,49].

Microscale measurements support vegetation as stabilizing and retaining aeolian accrued grains. In all collection types (seasonal, event, and monthly), transitional and unvegetated $\Delta z$ were largely equivalent, but one or both were larger than vegetated $\Delta z$ (Figure 4); despite this, the vegetated pins had the greatest net change (Figure 5) and in the seasonal and monthly collections, when both erosion and accretion occurred more concomitantly, vegetated pins were consistently associated with greater instances of accretion than erosion than would be expected by chance, whereas the opposite was true for the other pin types. Differences in vegetation density have been suggested as being a driver of topographic heterogeneity in foredune systems for over four decades [3,24,42,66–68]. However, there exist very few robust examples of topographic heterogeneity being directly attributable to biological factors in situ [34,37,38,69–71] relative to the wide breadth of research on how physical drivers impact transport and alter topography in these systems [57,72–74]. The data presented in this work is robust in both the temporal collection span and frequency and is consistent across analyses, such that we can be confident they are ecologically relevant and persistent [75]. It is important to note a lack of near-surface wind velocity measurements when discussing sediment transport drivers for the statistical findings. Wind velocities for used here to give an understanding of the overall mesoscale winds during seasonal, monthly, and event elevation pin collections. Nevertheless, these data (Table 4) can be a useful source of insight for modelers seeking to incorporate vegetative effects in ecogeomorphic simulation models towards better resource management and conservation efforts over varying spatiotemporal scales [5,56,57,72,76].

Beach access paths impacted localized elevation change and contributed to increased dynamism along path extents relative to vegetated dune areas [77,78]. This level of dynamism in $\Delta z$ was only matched at the dune crest itself and in blowouts (Figure 8) supporting the view that these areas as some of the most unstable and dynamic within the dune habitat, which in turn support the lowest plant species diversity and density [42,79,80]. The impact of access paths on dune systems’ ecology and geomorphology are both poorly understood and studied, but are likely self-reinforcing and will remain unvegetated provided they remain in use with both trampling and heightened abiotic forcings limiting seedling establishment and colonization [11,80–82] The largely linear unvegetated paths, here, provided longer fetch lengths when aligned with prevailing winds, thus enabling sediment movement inland from the foredune to the maritime forest and beyond [83,84]. This suggests that management efforts to alter or construct paths that purposefully include turns, especially out of phase with onshore winds (i.e., easterly at this site), could reduce the funneling of aeolian processes occurring during events and the subsequent migration of sand inland [85].

Comparisons of the two collection methods ultimately shed light on the pros and cons associated with each regarding what can and cannot be gleaned. These differences primarily stem from scale, microscale versus mesoscale, relative to the hypotheses in question [86]. Both metrics are snapshots capturing the state space of a dynamic system [87–89]. However, the remote sensing approach operates at the habitat or ecosystem levels, for both beaches and dunes, and appears better suited to explore geomorphological processes acting at both larger spatial and temporal scales [53,72]. Conversely, field pins appear better suited for exploring biological or ecological processes acting on smaller spatial and
temporal scales that are less clearly defined in this and other coastal ecogeomorphic systems [57,90]. When geological and biological processes are inextricably linked [9,91], the concomitant collection of both data types may yield more insight than the collection of one versus the other; this line of thinking has allowed other researchers in beach–dune habitats to test biological hypotheses that would previously have been difficult to explore without the use of both remote and field techniques [37,38,69,92].

The concomitant collection of both remote and field data can ultimately contextualize biological data to aid researchers in better understanding an observed or lack of observed pattern. For example, field collections of transect-level biological data may reveal no clear vegetation patterns where they might have been expected. However, zooming out with broader habitat-level topographic data might reveal that the system is in a state of recovery and self-reorganization towards a more stable state, wherein biological patterns may then begin to emerge and be observable [93–95]. The timing of the biological data collection may thus be off. As another example, Charbonneau et al. [38] did not observe the expected differences in the elevation of foredunes dominated by different species surrounding a native and invasive plant [96]. Conversely, Hacker et al. [69] and Goldstein and Moore [97] did observe species-specific topographic variation relative to ecogeomorphic feedbacks. This does not denote a lack of agreement within the system Charbonneau et al. studied; rather it is more likely that the relationships were undetectable when the data were collected because they had not been physically manifested yet, given the incongruent timescales on which biological and geomorphological feedbacks act ecogeomorphologically [9,87,91,98]. Event type, interval, and intensity for sites compound this sampling issue and must be considered when collecting and analyzing datasets [13,14].

Biological data collection can also contextualize geomorphic data towards understanding an observed or lack of observed pattern. Topographic data have proven very useful for evaluating the volumetric change of beach-dune systems pre- and post-storm when it may be most apparent [11–14,38]. Here, UAS surveying revealed broader landscape changes (Section 3.1) such as the system being accretional. However, collection frequency and vertical uncertainty were not precise enough to reflect the changes observed at microscale locations, overestimating Δz and not accurately predicting Δz as erosional or accretional (51–71% accuracy), thereby obscuring the vegetation-stabilizing effects apparent in all microscale collection timeframes; the high variability around the means between collections (Table 6) necessitates more rather than less data, regardless of data type (field or remote), to evaluate these ecogeomorphic feedbacks and suggests that changes in total elevation may only be visible after a critical period wherein sufficient events have occurred. In repeating this study, using additional ground control points within the survey area and the same UAS system, could have better controlled vertical uncertainty [14,53,99]. Increasing flight frequency can be a real challenge, given the need for appropriate weather and the associated costs (e.g., time, software, hardware, and data storage, processing, and extraction) relative to the benefits; this is especially true considering field measurements can be comparatively cheap, (e.g., physical effort, time, and data analysis) [14,100,101]. To collect remote data more frequently, one may need alternative methods than those presented here to reduce the costs associated with each survey, re-evaluate the hypothesis being posed in light of the data at hand, or foster interdisciplinary collaborations with the concomitant collection of supporting data.

5. Conclusions
Both the micro- and mesoscale metrics highlighted cause-and-effect relationships surrounding a beach–dune system’s topographic change, with both the identifying features of heightened dynamism for future monitoring and management efforts. Beach access paths, which funnel aeolian processes, should be counted among some of the most dynamic features in a dune habitat; mesoscale UAS topographic surfaces were necessary to quantify this, as well as habitat-level state changes such as overall beach-dune volumetric and geomorphological change. Plants’ presence can limit this dynamism as they played a
stabilizing role in the foredune, reducing sediment movement and capturing and eroding less in each accretional and erosional instance, respectively, than unvegetated and partially vegetated transition microhabitats. However, they better retained those grains, ultimately yielding the greatest net change in elevation over the two-year install of all pins, whereas erosion and accretion instances in the other microhabitats were largely ephemeral. Greater precision, accuracy, and data collection frequency of UAS surveys would be needed to adequately explore these plant-related ecogeomorphic feedbacks than was possible with mesoscale data. The cause-and-effect relationships gleaned from each data type and therein what is captured and missed, highlights the different spatial and temporal scales on which biological and geomorphological processes act in beach-dune systems. Bridging disciplines and data types may ultimately give researchers more power to tease apart complex interconnected ecogeomorphic relationships that structure dynamic beach-dune systems. These data, regardless of type and scale, are baseline snapshots that can ultimately help managers and practitioners to better understand where and how to target vulnerable areas to reduce growing coastal instability.

Supplementary Materials: The following are available online at www.mdpi.com/1660-4601/13/4/4488/s1. Supplementary 1: Field Pin Types. Supplementary 2: Beach and Dune Automated Feature Extraction. Supplementary 3: Beach Access Path Impacts Tool. Supplementary 4: Grain Size Distributions and Analyses. Supplementary 5: Meteorological Data and Associated Wind Roses.

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