Changes in Breath Trihalomethane Levels Resulting from Household Water-Use Activities

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Common household water-use activities such as showering, bathing, drinking, and washing clothes or dishes are potentially important contributors to individual exposure to trihalomethanes (THMs), the major class of disinfection by-products of water treated with chlorine. Previous studies have focused on showering or bathing activities. In this study, we selected 12 common water-use activities and determined which may lead to the greatest THM exposures and result in the greatest increase in the internal dose. Seven subjects performed the various water-use activities in two residences served by water utilities with relatively high and moderate total THM levels. To maintain a consistent exposure environment, the activities, exposure times, air exchange rates, water flows, water temperatures, and extraneous THM emissions to the indoor air were carefully controlled. Water, indoor air, blood, and exhaled-breath samples were collected during each exposure session for each activity, in accordance with a strict, well-defined protocol. Although showering (for 10 min) and bathing (for 14 min), as well as machine washing of clothes and opening mechanical dishwashers at the end of the cycle, resulted in substantial increases in indoor air chloroform concentrations, only showering and bathing caused significant increases in the breath chloroform levels. In the case of bromodichloromethane (BDCM), only bathing yielded a significantly higher air level in relation to the preexposure concentration. For chloroform from showering, strong correlations were observed for indoor air and exhaled breath, blood and exhaled breath, indoor air and blood, and tap water and blood. Only water and breath, and blood and breath were significantly associated for chloroform from bathing. For BDCM, significant correlations were obtained for blood and air, and blood and water from showering. Neither dibromochloromethane nor bromoform gave measurable breath concentrations for any of the activities investigated because of their much lower tap-water concentrations. Future studies will address the effects that changes in these common water-use activities may have on exposure. Key words: biomarkers, breath analysis, disinfection by-products, exposure, trihalomethane, water use. Environ Health Perspect 114:514–521 (2006). doi:10.1289/ehp.8171 available via http://dx.doi.org/ [Online 15 November 2005]
more acceptable to human subjects, resulting in higher levels of participation in exposure studies. Additionally, because the breath sample supply is virtually limitless, it allows the rapid collection of multiple samples and even lends itself to continuous real-time monitoring (Gordon et al. 1998). Because of the dynamic equilibrium between the concentration of a VOC in the blood and its concentration in exhaled breath (Wallace et al. 1996), breath measurements can be used to estimate body burden and to detect changes in body burden with time (Gordon et al. 1992; Wallace et al. 1993; Weisel et al. 1992).

The present study was undertaken to examine, through exhaled-breath measurements, which common household water-use activities lead to an increase in the internal dose levels of CHCl₃ and the other potentially harmful THMs in the human body. This will inform future studies by indicating which activities should be investigated more thoroughly. Water, air, blood, and breath samples were collected from subjects for each activity that was expected to elevate the internal dose of the THMs. An overview of all of the methods used in this study and a summary of the results obtained have been published by Nuckols et al. (2005). In this article, we show more clearly which activities are the major determinants of exhaled-breath THM concentrations, and focus on some of the correlations that are associated with these activities.

Materials and Methods

The experimental approach and methods used in this investigation have been presented elsewhere (Nuckols et al. 2005), so only a brief summary is given here. Emphasis is placed on those aspects that are particularly germane to this study.

Study locations and participants. The study was conducted in two single private homes, one in North Carolina (NC) and the other in Texas (TX). The selection of these locations was based upon the assumption that the distribution of THMs in the public water supplies that served these areas would be different because of anticipated high bromide concentrations in the TX raw water source. However, instead of the dominant bromine-containing species that were expected at the TX site, relatively low concentrations of BDCM, DBCM, and bromoform were found, probably due to heavy rains in the area during the week preceding the subject activities in the home, which diluted the bromide concentration in the raw water. Chloroform was the dominant THM species, and overall THM levels were less than expected.

Both homes were single-story, ranch-style residences (about 1,200 ft² or 111.5 m² total floor area) with three bedrooms, two bathrooms, and central heating, ventilating, and air conditioning (HVAC) systems. Tap water was provided in each case by the local public utility, which used combined chlorine as the residual disinfectant. The measured free chlorine residuals in the test homes at the time of sample collection were consistent with standard practice for chloraminated systems. THM levels at other locations in the distribution systems for the two utilities during the periods of this study were assumed to be similar because continuing THM formation in distribution systems is quenched in the presence of combined chlorine. The thermostat in each house was set at 75°F and the HVAC fan was set to the “on” position during the entire study period. The exhaust fan in the study bathroom was turned off throughout the study.

Seven healthy, young subjects, 21–30 years of age, were enrolled into the study: four in NC (three males and one female) and three in TX (one male and two females). Two of the males at the NC site reported their race as African American; all other participants reported their race as Caucasian. All subjects were nonsmokers and had a body mass index of 22–24. The study protocol was reviewed and approved by the institutional review boards of the Centers for Disease Control and Prevention (CDC) and the Battelle Memorial Institute; informed written consent was obtained from the subjects before they took part in the study.

Exposure regimen. To avoid inadvertent exposure to disinfection by-products, the subjects slept in the study residence the night before exposure activities began. Each subject took part in several common water-use activities over 2 distinct days that were roughly 1 week apart. The prescribed water-use activities are summarized in Table 1. Activities each day were separated by at least 1 hr (3 hr after showering and bathing) to ensure that blood and breath THM concentrations returned as close as possible to baseline levels before the start of the next exposure activity. Each activity was preceded and followed at fixed times by the collection of tap water and blood samples. To ensure privacy, each subject wore a bathing suit for the showering and bathing activities. Baseline indoor air and breath samples were collected at the start of each day before the subjects undertook any prescribed activities. Air samples then were collected during each activity, whereas breath samples were collected 5 min after each activity. Two of the subjects followed a more intensive protocol than the other subjects. These subjects performed the same activities as the other subjects, but additional breath, air, and water samples were collected during their shower activity. The additional samples were taken in an attempt to better define peak air and breath concentrations for the THMs during showering. The THM results obtained with the intensive shower protocol will be reported in a forthcoming article.

Sample collection and analysis. Shower and bath air. We collected air samples from the shower stall during each shower event to evaluate inhalation exposure for the subjects. We collected 13 samples for each of the five subjects who followed the standard protocol, and 19 samples for each of the two subjects who followed the intensive shower protocol. Depending on the specific water-use activity (Table 1), we manually collected either time-integrated or short-term (“grab”) air samples into precleaned and evacuated SUMMA 6-L stainless steel canisters (Scientific Instrumentation Specialists, Moscow, ID, and Biospherics, Hillsboro, OR). Grab samples were taken by rapidly opening the canister valve and allowing air to flow into the canister until near-atmospheric pressure equilibrium was attained (< 1 min). To collect time-integrated air samples, we attached precut stainless steel tubes of varying internal diameter and length to each canister. These tubes served as critical flow orifices that regulated the flow rate into the canister and ensured that it filled in approximately the same time as the time taken to complete the specific water-use activity being monitored. Integrated 13-min air samples associated with the showering activity were collected from the start of the showering activity until 3 min after the water was turned off. For the bathing activity, integrated 23-min air samples were collected from the start of the 6-min bath filling period, through the 14-min bathing (subject immersion) period, and ended 3 min after the subject got out of the bathtub.

For the standard shower protocol, the single canister used to collect shower stall air

| Day | Water-use activity |
|-----|-------------------|
| 1 | 1. Baseline measurements before any water use |
| 2 | 2. Drink cold water-based (fruit juice) beverage (0.25 L) |
| 3 | 3. Hot-water shower^ |
| 4 | 4. Drink cold tap water (0.5 L) |
| 5 | 5. Automatic clothes washing^ |
| 6 | 6. Hand washing |
| 7 | 7. Automatic dishwashing |
| 8 | 8. Open/remove dishes from dishwasher at end of cycle |
| 9 | 9. Open/remove clothes from dryer at end of wash cycle for washing machine |
| 10 | 10. Open/remove clothes from dryer at end of wash cycle for washing machine |
| 11 | 11. Automatic clothes washing; bleach added during wash cycle^ |
| 12 | 12. Hand washing of dishes |
| 13 | 13. Subject stayed in room adjoining study bathroom |
| 14 | 14. Subject stayed in room adjoining study bathroom |
| 15 | 15. Subject stayed in room adjoining study bathroom |
| 16 | 16. Subject stayed in room adjoining study bathroom |
| 17 | 17. Subject stayed in room adjoining study bathroom |
| 18 | 18. Subject stayed in room adjoining study bathroom |
| 19 | 19. Subject stayed in room adjoining study bathroom |
| 20 | 20. Subject stayed in room adjoining study bathroom |
| 21 | 21. Subject stayed in room adjoining study bathroom |
| 22 | 22. Subject stayed in room adjoining study bathroom |
| 23 | 23. Subject stayed in room adjoining study bathroom |

^Five subjects followed the standard shower protocol; two subjects undertook the intensive shower protocol. | participant did not stay in same room while washing machine was operating. | Bathroom door was opened at end of shower event.
was suspended from the shower curtain rod close to the shower head; for the bath activity, the canister was placed on the bathtub ledge close to the subject’s head. Because of space limitations in the bathrooms and the need to rapidly collect multiple canister grab samples (~1–2 min apart) for the intensive shower protocol, three specially constructed stainless steel manifolds were used, one to sample the bathroom air, one to collect the shower stall air, and the third to collect the breath samples (Figure 1). One end of the sampling tube for bathroom air was positioned in the bathroom, and sampling was performed remotely in an adjoining room via a length of stainless steel tubing (0.22 cm i.d., ~6.0 m long) attached to the manifold. For the collection of air and breath samples in the shower stall, two separate lengths of stainless steel tubing (0.22 cm i.d., ~16 m long) were used. The tube used to collect breath samples was connected to a T-piece that joined the face mask and breath containment coil; the other end was connected to the remote breath sampling manifold located outside the bathroom. Sampling with all three systems involved manual opening of each canister valve in turn and then closing the valve at the end of the sampling period.

The breath containment coil (899 cm × 1.3 cm i.d.), which is based on the alveolar breath sampling device developed by Raymer et al. (1990), provides a sufficiently large buffer volume (1.1 L) that ensures that the breath being sampled is primarily alveolar because it is from the end of the expiration. The canisters were stored at room temperature and, at the end of each sampling day, shipped by overnight express courier to the Atmospheric Science and Applied Technology Laboratory at Battelle (Columbus, OH) for THM analysis.

Samples were analyzed for THMs by automated gas chromatography/mass spectrometry (GC/MS) using a modified version of U.S. Environmental Protection Agency (EPA) Method TO-14 (Winberry et al. 1990). The GC was connected to a cryogenic preconcentration trap, which was cooled to −185°C for sample collection and heated to 120°C during sample desorption. A six-port valve controlled sample collection and injection. The preconcentrator was also equipped with an autosampler so that up to 16 canister samples could be analyzed automatically. The sample volume was 90 mL. Analytes were chromatographically resolved on a Restek RTX-1 fused silica capillary column (Restek Corp., Bellefonte, PA). The mass spectrometer was operated in the full scan mode. Target analytes were identified by matching the mass spectra acquired from the sample to the mass spectral library from the National Institute of Standards and Technology (Gaithersburg, MD). Quantification of all identified peaks was based upon multipoint calibration curves, which were generated for each target analyte at the start of the study. During each analysis period, a single-point calibration was run; precision for each THM was typically <20% relative SD.

**Exhaled breath.** As shown in Table 1, we monitored eight water-use activities during the subject’s day-1 sampling visit and seven water-use activities during the day-2 visit, approximately 1 week later. We collected a total of 15 breath samples from each of the five subjects who participated in the standard shower protocol, and 22 samples from each of the two subjects who followed the intensive shower protocol. In all cases, except while the subject was in the room adjacent to the operating shower (i.e., secondary shower exposure activity), breath samples were collected 5 min after completion of the activity. The single breath canister (SBC) procedure, developed by Pleil and Lindstrom (1995a), was used to collect the breath samples, except in the case of the intensive shower protocol. The SBC sampling method is a self-administered procedure in which the subject exhales alveolar air directly into an evacuated 1-L Silcosteel stainless steel canister (Entech, Simi Valley, CA) fitted with a short Teflon tube that serves as a disposable mouthpiece. To obtain a breath sample, the subject covers his or her lips over the open end of the Teflon tube while exhaling, opens the canister valve, and fills the evacuated volume. The subject is instructed to begin sample collection near the end of a normal resting tidal breath to provide what is mostly alveolar breath.

For the intensive shower protocol, the subject was fitted with a face mask (model 8932; Hans Rudolph, Inc., Kansas City, MO) that covered his or her mouth and nose and was equipped with a two-way non-rebreathing valve set. The subject inhaled through the one-way valve in the inhalation port, which was left open to the shower stall air. The exhalation port of the mask was attached via a stainless steel tube to a convoluted polytetrafluoroethylene breath containment coil in an adjoining room (Figure 1). The subject exhaled normally through the one-way valve in the exhalation port. Exhaled breath (primarily alveolar) was collected remotely in an evacuated 1-L Silcosteel canister, attached to the breath sampling manifold by manually opening the canister valve and allowing the breath to flow back through the containment coil, back through the stainless steel sampling line, and into the canister until the contents of the canister reached near-atmospheric pressure (≤1 min). As in the case of the intensive bathroom air samples, the sampling manifold allowed the rapid collection of successive grab breath samples.

Background samples were obtained once each morning before any water-use activities began. Samples were shipped at the end of the day by overnight express courier to Battelle for THM analysis, which was carried out by automated GC/MS as described above. To correct for dilution of alveolar air from dead volume air in the breath sample and to normalize the concentration of each analyte to correctly...
reflect the breath concentration, we measured the carbon dioxide concentration in each canis-
ter sample using a CO2 monitor (Pryon model SC-300; Pryon Corp., Menomonee Falls, WI)
equipped with an external infrared CO2 sen-
sor. Analysis of CO2 in the breath provides an
accurate correction factor for the approximate
amount of alveolar breath (as opposed to
whole breath) collected in the sample (Pleil
and Lindstrom 1995b).

Blood. Each participant provided a total of
26 blood samples over the course of the 2-day
study; 14 on day 1 and 12 on day 2. Samples
were taken approximately 5 min before and
after each activity (and 30 min after the end of
the shower and bath activities), using a venous
catheter that remained in the subject’s arm
throughout each study day (~12 hr). Blood
collection (Vacutainer) tubes were specially
treated before use to remove background cont-
amination (Cardinali et al. 1995). After collect-
ion, samples were refrigerated and then
packed into coolers with ice packs for shipping
by overnight express courier to the Division of
Laboratory Sciences of the CDC (Atlanta, GA)
for analysis. Details of these procedures have
been described elsewhere (Bonin et al. 2005).

The whole blood samples were analyzed for
THMs using solid-phase microextraction
GC/isotope dilution MS, with the mass spec-
trometer operating in the selected ion
monitoring mode. Stable isotopically labeled
analogs of the compounds of interest were
added to the samples as internal standards,
and quantification was accomplished by meas-
uring specific ion responses relative to those
of the corresponding labeled analogs (Ashley
et al. 1992; Cardinali et al. 2000).

Water. We collected a total of 21 (normal
shower protocol) or 22 (intensive shower pro-
tocol) water samples over each 2-day study
period. All water samples were collected in
headspace-free 40-mL acid-washed glass vials
with screw caps. Residual chlorine was
quenched with ammonium sulfate. Besides
the samples of water associated with each
activity, we collected several samples from a
cold water tap during each exposure day to
establish baseline THM concentrations. The
temperature of the water used in each activity
was controlled over a limited range and
recorded as samples were collected for each

| Location | Water concentration (µg/L) for CHCl₃ and BDCM over all subjects by exposure day and location. |
|----------|------------------------------------------------------------------------------------------|
|          | Water concentration (mean ± SD)             | CHCl₃          | BDCM          |
| NC       | 1  128.2 ± 14.4  31.9 ± 11.4             | 90.4 ± 6.4     | 24.4 ± 7.7   |
| TX       | 1  28.5 ± 4.1    10.6 ± 2.8              | 22.2 ± 2.5     | 8.6 ± 2.0    |

Results

Effects of tap-water quality and water-use
activity on indoor air and exhaled-breath
congenations. Table 2 summarizes the over-
all mean water concentrations for CHCl₃ and
BDCM over all subjects by exposure day and
sampling site (NC and TX). A detailed analy-
sis of the data indicates that there were large
changes in the THM concentrations over the
course of the exposure measurements
(Nuckols et al. 2005). Mean water concentra-
tions were much higher on both exposure
days at the NC site than in TX for CHCl₃ as
well as for BDCM.

The effect of water-use activities on the
mean indoor air and exhaled-breath concen-
trations of CHCl₃ normalized with respect to
tap-water concentrations (micrograms per
cubic meter of breath per microgram per liter
of water) for all subjects on both exposure
days and at both sites is shown in Figure 2.
Most indoor air samples for the water-use
activities shown on the x-axis in Figure 2A
were taken during the period of each exposure
activity, whereas most exhaled-breath samples
shown in Figure 2B were collected 5 min
after each exposure activity ended. Compared
with the relevant day-1 or day-2 preexposure
(baseline) level, the increase in indoor air con-
centration of CHCl₃ was greatest for subjects
as a group at both sites for the showering
and the secondary shower exposure activities
(>40-fold increase), followed by the bathing
activity (>10-fold increase), and post-
secondary shower exposure and exposure due
to the machine washing of clothes using
bleach (>5-fold increase).

Figure 2B shows the mean normalized
exhaled-breath concentrations of CHCl₃ for
the preexposure samples and those collected
5 min after each activity ended. The exhaled-
breath levels are approximately an order of
magnitude lower than the levels in the indoor
air for the corresponding activities. The activi-
ties that resulted in marked increases in breath
levels over the corresponding baseline level
across all subjects at both sites were showering
(5-fold increase) and bathing (6-fold increase).
The stronger increases in relative air concentra-
tions noted above for the secondary shower
exposure, the postsecondary shower exposure,
and exposure due to the mechanical washing of
clothes using bleach were not reflected in the
corresponding breath measurements.

The mean normalized indoor air and
exhaled-breath concentrations of BDCM for
the preexposure samples and defined activity
samples are shown in Figure 3. Although the
concentrations of BDCM in the tap water,
and hence in the air and breath samples, were
much lower than those for CHCl₃, these plots
are generally very similar to those obtained for
CHCl₃, but with greater variability in the
data. Although the air BDCM levels for the

Table 2. Overall mean water concentrations (µg/L) for CHCl₃ and BDCM over all subjects by exposure day and location.

| Location | Exposure day | CHCl₃ (mean ± SD) | BDCM (mean ± SD) |
|----------|-------------|------------------|------------------|
| NC       | 1           | 128.2 ± 14.4     | 90.4 ± 6.4       |
|          | 2           | 90.4 ± 6.4       | 24.4 ± 7.7       |
| TX       | 1           | 28.5 ± 4.1       | 22.2 ± 2.5       |
|          | 2           | 10.6 ± 2.8       | 8.6 ± 2.0        |
showering and secondary shower exposure activities are much greater than their baseline levels (> 10-fold increase in each case), followed by the bathing activity (4-fold increase), only the showering and bathing activities are moderately higher in the breath BDCM measurements (< 2-fold increase in each case).

Correlations among breath, blood, water, and air measurements. Table 3 presents Spearman correlation coefficients for CHCl₃ and BDCM for blood and breath in samples obtained from the showering and bathing activities undertaken by all subjects. For CHCl₃ from the shower event, Spearman coefficients are highly significant for breath and air, blood and breath, air and blood, and water and blood. For CHCl₃ from the bathing event, only the water-breath and blood-air correlations are significant. For BDCM, the significant correlations are those for blood and air and for blood and water from the shower activity.

The data presented in Table 4 suggest that the mean air and exhaled-breath concentrations, normalized with respect to the water concentrations, were different for the showering and bathing activities. For both CHCl₃ and BDCM, the normalized mean air concentrations were much greater for the showering activity than for the bathing activity. In contrast, the breath concentrations for the showering and bathing activities were significantly smaller and essentially the same for the two activities. The ratios of the normalized mean breath to air concentrations were, however, much greater for the bathing activity than for the showering activity (0.37 vs. 0.08 for CHCl₃; 0.40 vs. 0.11 for BDCM).

Figure 4 shows plots of the CHCl₃ concentrations in exhaled breath taken 5 min after the bath or shower exposure ended versus the air concentrations measured during the exposure period. The breath concentrations clearly increase with increasing air concentrations for these two events, and the slopes of the plots are quite similar, with the breath levels for the bathing activity greater than those for the showering activity. Figure 5 presents the CHCl₃ concentrations in exhaled breath versus the corresponding blood concentrations, both of which were taken 5 min after the bath or shower exposure ended. The breath concentrations and blood concentrations correlate closely for the two activities.

Discussion

Effect of common water-use activities on indoor air and exhaled-breath concentrations. The primary purpose of this study was to examine which common household water-use activities result in an increase in exhaled-breath concentrations and, hence, the internal dose of THMs in people conducting these activities. The results presented in Figure 2 for CHCl₃ show that the showering and secondary shower exposure activities, as well as bathing, postsecondary shower exposure, and exposure due to mechanical washing of clothes with bleach resulted in appreciable increases in air concentrations for all subjects at both sampling sites, despite the large difference in tap-water CHCl₃ concentration between the two sites.

Although much lower than the corresponding air levels, the mean normalized postexposure breath concentrations of CHCl₃...
were consistently higher than the initial pre-exposure levels for the 10-min showering activity (mean ± SE, 0.19 ± 0.02 vs. 0.04 ± 0.01 µg/m³ per µg/L) and the 14-min bathing activity (0.25 ± 0.04 vs. 0.07 ± 0.03 µg/m³ per µg/L). Despite lower post-exposure air and breath concentrations of BDCM than those obtained for CHCl₃, the air concentrations associated with the showering, secondary shower exposure, and bathing activities were markedly higher than their pre-exposure concentrations. Only the showering and bathing activities resulted in moderately higher post-exposure breath concentrations of BDCM. As a result of the much lower DBCM and bromoform concentrations in the tap water in this study, neither chemical gave measurable breath concentrations for any of the water-use activities investigated (Nuckols et al. 2005).

There have been numerous laboratory-based studies of controlled exposure to THMs, especially CHCl₃, from showering, bathing, or other household activities such as using a dishwasher (Andelman 1985; Gordon et al. 1998; Jo et al. 1999a, 1999b; Olson and Corsi 2004; Weisel and Jo 1996; Weisel et al. 1992). Those few that have been conducted in normal residential settings have been restricted to one or two common water-use activities and to measurements of CHCl₃ concentrations in indoor air or exhaled breath. Table 5 shows that our observed normalized indoor air and exhaled-breathe levels of CHCl₃ and BDCM as a result of showering are consistent with results reported in previous residential exposure studies (Egorov et al. 2003; Kerger et al. 2000; May et al. 1995). Although the tap-water concentrations during showering vary markedly across the four studies cited in Table 5 (CHCl₃, 47–198 µg/L; BDCM, 7–42 µg/L), the mean normalized indoor air concentrations during showering are quite similar (CHCl₃, 1.67–3.52 µg/m³ per µg/L; BDCM, 1.25–1.91 µg/m³ per µg/L). The simple normalization procedure applied in the present study, and to the data from the three previous residential studies summarized in Table 5 (Egorov et al. 2003; Kerger et al. 2000; May et al. 1995), indicates generally good agreement between the studies. It should, however, be noted that in our study, the shower exposure conditions, and therefore the extent of volatilization, were uniform for all subjects.

The release of CHCl₃ and BDCM during showering activities, expressed in terms of the mean normalized air concentrations in Table 5, shows a consistent concentration gradient of CHCl₃ > BDCM. From our data, the mean normalized air concentration ratio (Cₐir/Cwater) for CHCl₃:BDCM is 1.17, which is within the range of values obtained for the other residential studies listed in Table 5 (1.12–1.84). To determine whether our measured concentration ratio is consistent with the physical properties of these compounds, we used a simple equilibrium model for predicting concentrations of VOCs in shower stall air (Sanders 2002). In this model, the entire mass of each chemical entering the system during the shower event is assumed to partition between the water and air phases under equilibrium conditions; loss of compound from the shower system by air exchange is ignored. All of the VOCs transferred from the water to the air phase during the showering event are assumed to remain in the shower stall. From the model, which is expressed in terms of a temperature-adjusted, dimensionless Henry’s law constant, H, and site-specific parameters, the ratio of the normalized concentrations of CHCl₃ and BDCM in the air compartment at equilibrium is given by

\[
\frac{C_{\text{air}}}{C_{\text{water}}}, \text{CHCl}_3 = \frac{H_{\text{CHCl}_3} \times V_{\text{air}} + V_{\text{water}}}{H_{\text{BDCM}} \times V_{\text{air}} + V_{\text{water}}}\]

where \(V_{\text{air}}\) is the volume of the shower stall, and \(V_{\text{water}}\) is the total volume of water used during a shower event.

We used average values for the site-specific parameters, namely, 1.74 m³ for the volume of the shower stall, 0.08 m³ for the volume of the shower water (10-min shower duration with 8.0 L/min flow rate), no air exchange, and a shower water temperature of 41°C for calculating the dimensionless Henry’s law constants (Sander 1999; U.S. EPA 2005) for CHCl₃ and BDCM. From Equation 1, the calculated normalized air concentration ratio for CHCl₃:BDCM is 1.166, in excellent agreement with the experimentally measured ratio. This suggests that the simple equilibrium model may be used to predict the concentration ratios of THMs in

### Table 3. Spearman rank correlations among air and water exposure measurements and breath and blood biomarkers in shower and bath samples from all subjects.

| Analyte (medium) | Air Breath Blood | Air Breath Blood |
|------------------|------------------|------------------|
| CHCl₃ (air)      | 0.94*            | 0.99**           |
| CHCl₃ (water)    | 0.83             | 0.61             |
| CHCl₃ (blood)    | 0.83             | 0.75             |
| BDCM (air)       | 0.94*            | 0.93**           |
| BDCM (water)     | 0.54             | 0.54             |
| BDCM (blood)     | 0.54             | 0.71             |
| CHCl₃ (air)      | 0.43             | 0.14             |
| CHCl₃ (water)    | 0.43             | 0.14             |
| CHCl₃ (blood)    | 0.43             | 0.14             |

*p < 0.05; **p < 0.005.

### Table 4. Normalized mean (± SE) air and exhaled-breath concentrations (µg/m³ per µg/L) for CHCl₃ and BDCM over all subjects from showering and bathing activities.

| THM component | Air | Breath |
|---------------|-----|--------|
| CHCl₃         | 2.28 ± 0.20 | 0.68 ± 0.12 |
| BDCM          | 2.09 ± 0.27 | 0.55 ± 0.10 |
shower systems and obtain insights into the behavior of those brominated THMs for which we have little or no data.

Published reports on the measurement of THM breath concentrations from exposures in residential settings are sparse. Table 5 includes our mean normalized breath concentration data for CHCl₃ and BDCM along with values reported recently by Egorov et al. (2003). Although both data sets show similar concentration gradients from CHCl₃ to BDCM as was found for shower air, our CHCl₃ and BDCM breath concentrations and the CHCl₃/BDCM ratio are markedly lower than the values obtained by Egorov et al. (2003). Two important reasons for these differences are the shower duration and the time when the samples were taken relative to the end of the exposure period. In our study, participants showered for 10 min, whereas participants in the Egorov et al. (2003) study showered for 15–20 min. The shorter showering period in our study would have resulted in a decrease in the air levels by as much as 50%. Additionally, previous work has shown that postexposure exhaled-breath concentrations of CHCl₃ decrease rapidly with time, especially in the first few minutes after exposure (Gordon et al. 1998). The uptake and elimination residence times for BDCM are not known and are unlikely to be the same as those for CHCl₃. Because breath samples were taken a full 5 min after exposure ceased in our study, whereas breath samples were collected within 1 min after the subjects completed their showering activity in the Egorov et al. (2003) study, differences in these factors may explain the large differences in the observed breath values in Table 5.

Other differences that may limit the comparability between studies in Table 5 include the presence of a minimal free chlorine residual and the lack of control over shower duration, flow rate, and temperature among the subjects from different households in the Egorov et al. (2003) study.

Relationships between air, water, blood, and breath measurements. We examined the relationships between the concentrations of CHCl₃ and BDCM associated with both the showering and bathing activities in water, air, blood, and breath. The data in Table 4 indicate that the mean normalized air concentrations of CHCl₃ and BDCM are much lower for the bathing event, probably due to the lesser opportunity for gas transfer between the bath water and the air. In contrast, the higher volatilization rate of CHCl₃ and BDCM into the shower stall air from the shower spray was the likely reason for the significantly higher air levels observed for the showering activity.

Despite the higher CHCl₃ and BDCM air concentrations noted in Table 4 for the showering activity than for the bathing activity, the respective breath concentrations are essentially the same. This suggests that the overall rate of uptake of the two THM species is roughly the same and that the dermal route is the more important exposure pathway (vs. inhalation) in the bath than in the shower. Further confirmation of the importance of the dermal route is provided by a comparison of the effect of secondary shower exposure on the measured air and breath concentrations (Figure 2). The large increase observed in the air concentration with secondary shower exposure almost matches, not surprisingly, that for the showering activity. However, contrary to the role of the showering activity in raising the breath concentration, the elevated air concentration due to the secondary shower exposure had no such effect on the breath concentration.

The relatively strong correlations between blood CHCl₃ and breath CHCl₃ concentrations suggested by the Spearman correlation coefficients for the showering and bathing activities are further supported by the plots in Figure 5. The association derives from the dynamic equilibrium that is assumed to exist between the concentration of the volatile compound in arterial blood and its concentration in alveolar breath (Wallace et al. 1996). Thus, the breath concentration can be used as a proxy to estimate body burden or changes in body burden with time (Gordon et al. 1992; Raymer et al. 1991; Wallace et al. 1993; Weisel et al. 1992). The slopes of the regression lines in Figure 5 may be used to estimate the average venous blood-to-breath ratio for CHCl₃. The average CHCl₃ blood-to-breath ratios from the showering and bathing activities were estimated to be 6.9 ($R^2 = 0.84$) and 7.5 ($R^2 = 0.61$), respectively, which compare favorably with the blood/air partition coefficient (6.9) reported by Gargas et al. (1989), but are smaller than the value (10.7) reported more recently by both Fisher et al. (1997) and Batterman et al. (2002).

In summary, we found that showering and bathing are two common household water-use activities that cause significant increases in exhaled breath concentrations of CHCl₃. Further analysis of the data for CHCl₃ from showering indicated strong correlations for indoor air and exhaled breath, blood and exhaled breath, indoor air and blood, and tap water and blood. For CHCl₃ from bathing, only water and breath, and blood and breath showed strong associations. Future studies are warranted to explore changes in these activities that can affect exposure.

### Table 5. Comparison of normalized CHCl₃ and BDCM concentrations in tap water, during-shower air, and postshower breath from THM exposure studies.

| Target matrix | CHCl₃ | BDCM |
|--------------|------|------|
| Mean water concentration (µg/L) | | |
| May et al. (1995) | 55 | 17 |
| Kerger et al. (2000) | 47 | 42 |
| Egorov et al. (2003) | 198 | 6.7 |
| This study | 82.8 | 20.3 |
| Mean normalized during-shower air concentration (µg/m³ per µg/L) | | |
| May et al. (1995) | 1.72 | 1.54 |
| Kerger et al. (2000) | 3.52 | 1.91 |
| Egorov et al. (2003) | 1.67 | 1.25 |
| This study | 2.23 ± 0.18 | 1.91 ± 0.23 |
| Mean normalized postshower breath concentration (µg/m³ per µg/L) | | |
| May et al. (1995) | NA | NA |
| Kerger et al. (2000) | 0.54 | 0.12 |
| Egorov et al. (2003) | 0.15 ± 0.01 | 0.07 ± 0.02 |

NA, not analyzed.

*In water source, bromoform was near or below LOD at most sites; in air samples, DBCM and bromoform were < LODs in Egorov et al. (2003) study and this study. n = 44 in May et al. (1995) study; n = 20 for source water and n = 12 for shower air in Kerger et al. (2000) study; n = 14 for source water, n = 35 for shower air, and n = 9 for exhaled breath in Egorov et al. (2003) study; and n = 6 for source water, shower air, and exhaled breath in this study. *Data for mean values for CHCl₃ and BDCM for source water and shower air estimated from plots in May et al. (1995). *Air concentration data obtained by Kerger et al. (2000) from eight unventilated and four vented shower events. In vented events, either the bathroom exhaust fan was on or the bathroom window was opened during the sampling event. The bathroom door was shut for all shower sampling events. *Shower duration: May et al. (1995) 10 min; Kerger et al. (2000) 8.6 min and 12 min; Egorov et al. (2003) 15–20 min; this study, 10 min. *Breath sample collection: Egorov et al. (2003) ≤ 1 min postexposure; this study, 5 min postexposure.

### References

Andelman JB. 1985. Inhalation exposure in the home to volatile organic contaminants of drinking water. Sci Total Environ 47:443–460.

Ashley DL, Bonin MA, Cardinal FL, McCraw JM, Holler JS, Needham LL et al. 1992. Determining volatile organic compounds in human blood from a large sample population by using purge and trap gas chromatography/mass spectrometry. Anal Chem 64:1021–1028.

Ashley DL, Prab JD. 1997. Time dependence of blood concentra- tions during and after exposure to a mixture of volatile organic compounds. Arch Environ Health 52:36–33.

Backer LC, Ashley DL, Bonin MA, Cardinal FL, Kieszak SM, Wooten JV. 2000. Household exposures to drinking water disinfection by-products: whole blood trihalomethane levels. J Expo Anal Environ Epidemiol 10:231–326.

Batterman S, Zhang L, Wang S, Fransblau A. 2002. Partition
coefficients for the trihalomethanes among blood, urine, water, milk, and air. Sci Total Environ 284:237–247.
Benoit FM, Nicolodakis H, Cardinal C, Alleyn C, Mori B. 1999. Characterization of human exposure to chloroform in an experimental shower by breath analysis (Abstract). In: Proceedings of the 46th ASMS Conference on Mass Spectrometry and Allied Topics, 31 May–4 June 1998, Orlando, FL. Santa Fe, NM:American Society for Mass Spectrometry, 74D.
Bonin MA, Silva L, Smith MM, Ashley DL, Blount BC. 2005. Measurement of trihalomethanes and methyl tert-butyl ether in whole blood using gas chromatography with high resolution mass spectrometry. J Anal Toxicol 29:61–69.
Bowe F, Fulcomer MC, Savin JE. 1995. Public drinking water contamination and birth outcomes. Am J Epidemiol 141:850–862.
Cantor KP, Hoover R, Hartge P, Mason TJ, Silverman DT, Altman R, et al. 1987. Bladder cancer, drinking water sources, and tap water consumption: a case-control study. J Natl Cancer Inst 79:1269–1279.
Cardon FL, Ashley DL, Bonin M, Wooten J. 1995. Treatment of Vacutainers for use in the analysis of volatile organic compounds in human blood at the low parts-per-trillion level. J Chromatogr Sci 33:557–560.
Churchill JE, Ashley DL, Kaye WE. 2001. Recent chemical exposures and blood volatile organic compounds in a large population-based sample. Arch Environ Health 56:157–166.
Egorov AI, Tereschenko AA, Altshul LM, Vartiainen T, Samsonov D, LaBrecque B, et al. 2003. Exposures to drinking water chlorination-by-products in a Russian city. Int J Hyg Environ Health 206:539–551.
Fisher J, Mahle D, Bankston L, Greene R, Gearhart J. 1997. Lactational transfer of volatile chemicals in breast milk. Am Ind Hyg Assoc J 58:425–431.
Gargas ML, Burgess RJ, Vosder CE, Carson GH, Andersen ME. 1989. Partition coefficients of low-molecular-weight volatile chemicals in various liquids and tissues. Toxicol Appl Pharmacol 98:71–79.
Gordon SM, Kenny DJ, Kelly T. 1992. Continuous real time breath analysis for the measurement of half lives of expired volatile organic compounds. J Expo Anal Environ Epidemiol 10:581–585.
Kerger BD, Schmidt CE, Faustenbach D.J. 2000. Assessment of airborne exposure to trihalomethanes from tap water in residential showerers and baths. Risk Anal 20:637–651.
Levesque B, Ayotte P, Tardif R, Ferron L, Greggs S, Schloch E, et al. 2008. Cancer risk associated with household exposure to chloroform. J Toxicol Environ Health A 65:488–502.
Lynberg M, Nuckols JR, Langlois P, Ashley D, Singer P, Mendola P, et al. 2001. Assessing exposure to disinfection by-products in women of reproductive age living in Corpus Christi, Texas, and Cobb County, Georgia: descriptive results and methods. Environ Health Perspect 109:597–604.
May SJ, Breuer GM, Kestle PA. 1995. The Determination and Comparison of Trihalomethanes in Water and Air in Homes Using Disinfected Surface Water, Disinfected Ground Water, and Non-chlorinated Well Water. Final Report. Iowa City, IA:Center for Health Effects of Environmental Contamination.
Miles AM, Singer PC, Ashley DL, Lynberg M, Mendola P, Langlois PH, et al. 2002. Comparison of trihalomethanes in tap water and blood. Environ Sci Technol 36:1962–1968.
Morris RD, Auert AM, Angelillo IF, Chalmers TC, Mosteller F. 1992. Chlorination, chlorination-by-products, and cancer: a meta-analysis. Am J Public Health 82:953–963.
Nieuwenhuisen JJ, Toledano MB, Elliott P. 2000. Uptake of chloroform by-products, a review and a discussion of its implications for exposure assessment in epidemiological studies. J Expo Anal Environ Epidemiol 10:586–599.
Pelou JD, Lindstrom AB. 1995a. Collection of a single alveolar exhaled breath for volatile organic compounds analysis. Am J Ind Med 27:109–113.
Pelou JD, Lindstrom AB. 1995b. Measurement of volatile organic compounds in exhaled breath as collected in evacuated electropolished canisters. J Chromatogr B 665:271–278.
Raymer JH, Pellizzari ED, Thomas KW, Cooper SD. 1991. Elimination of volatile organic compounds in breath after exposure to occupational and environmental micro-environments. J Expo Anal Environ Epidemiol 1:438–451.
Raymer JH, Thomas KW, Cooper SD, Whitaker DA, Pellizzari ED. 1990. A device for sampling of human alveolar breath for the measurement of expired volatile organic compounds. J Anal Toxicol 14:337–344.
Sander R. 1999. Compilation of Henry’s Law Constants for Inorganic and Organic Species of Potential Importance in Environmental Chemistry (Version 3). Available: http://www.mpch-mainz.mpg.de/~sander/res/henry.html [accessed 1 October 2004].
Sander PF. 2002. A Screening Model for Predicting Concentrations of Volatile Organic Chemicals in Shower Stall Air. Available: http://www.state.nj.us/dep/dsr/ research/that/2bohwers%20paper%20.pdf [accessed 21 August 2005].
Singer PC, DePaz E, Ashley DL, Blount B, Nuckols JR, Wilkes CR, et al. 2003. Impact of trihalomethane concentrations in tap water and water use activities on biological levels of trihalomethanes. In: Proceedings of the Water Quality Technology Conference, American Water Works Association, 2–6 November 2003, Philadelphia, PA, Denver, CO:American Water Works Association.
U.S. EPA. 1995. Determination of Chlorination Disinfection By-products, Chlorinated Solvents, and Halogenated Pesticides/Herbicides in Drinking Water by Liquid-Liquid Extraction and Gas Chromatography with Electron-Capture Detection. Method 551.1, Cincinnati, OH:National Exposure Research Laboratory, Office of Research and Development, U.S. Environmental Protection Agency.
U.S. EPA (Environmental Protection Agency). 2005. Estimated Henry’s Law Constant. Available: http://www.epa.gov/ ncbetns/learn2model/part-two/onsite/esthenry.htm [accessed 8 September 2005].
Vena JE, Graham S, Freundheim J, Marshall J, Zielonyz M, Swanson M, et al. 1993. Drinking water, fluid intake, and bladder cancer in western New York. Arch Environ Health 48:191–198.
Wallace LA. 1997. The TEAM Study: Summary and Analysis: Volume I. EPA/600/6-87/002a. Washington, DC:U.S. Environmental Protection Agency.
Wallace LA. 1997. Human exposure and body burden for chloroform and other trihalomethanes. Crit Rev Environ Sci Technol 27:113–194.
Wallace LA, Bickley T, Pellizzari E, Gordon SM. 1996. Breath measurements as VOC biomarkers. Environ Health Perspect 104(suppl 5):861–869.
Wallace LA, Pellizzari ED, Gordon SM. 1993. A linear model relating breath concentrations to environmental exposures: application to a chamber study of four volunteers exposed to volatile organic chemicals. J Expo Anal Environ Epidemiol 3:75–102.
Weisel CP, Jo WK. 1996. Ingestion, inhalation, and dermal exposure to chloroform and trichloroethene from tap water. Environ Health Perspect 104:48–51.
Weisel CP, Jo WK, Loya PJ. 1992. Utilization of breath analysis for exposure and dose estimates of chloroform. J Expo Anal Environ Epidemiol Suppl 1:59–70.
Weisel CP, Kim H, Haltereier P, Klotz JA. 1999. Exposure estimates to disinfection by-products of chlorinated drinking water. Environ Health Perspect 107:103–110.
Winbom WT Jr, Murphy NT, Riggin RM. 1990. Methods for Determination of Toxic Organic Compounds in Air: EPA Methods. Park Ridge, NJ:Noyes Data Corporation.