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Hierarchically structural layered double oxides with stretchable nanopores for highly effective removal of protein-bound uremic toxins

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ABSTRACT

The global outbreak and prevalence of coronavirus disease 2019 (COVID-19) has triggered an urgent demand for family hemodialysis equipment. It is particularly vital to design and apply superior adsorbents to adsorb toxins for reducing the usage of dialysate. In this work, hierarchically structural MgAl layered double oxides (LDO) with stretchable nanopores were exploited through a facile one-pot trisodium citrate (TSC) assistant hydrothermal reaction followed by calcination treatment for effectively adsorbing protein-bound uremic toxins such as hippuric acid (HA) or indoxyl sulfate (IS). The optimized MgAl LDO possessed flower-like spherical morphology, ultrahigh specific surface area (187.3 m$^2$/g) and uniquely stretchable nanopores, which were more conducive to incorporating anions due to their unique memory effect endowing them with promising adsorption capacities for HA or IS. And the adsorption data could be better conformed to pseudo-second-order kinetic model and Langmuir isotherm determining that the maximum adsorption capacity of HA and IS was 129.8 mg/g and 63.1 mg/g, respectively. Furthermore, the computation of molecular size paired with the analysis of adsorption mechanism accurately revealed that high-efficiency toxin capture was mainly attributed to electrostatic interaction for internal intercalation and surface adsorption. Therefore, the application of such delicate LDO as new premium adsorbent would facilitate the development and popularization of family hemodialysis equipment.

1. Introduction

Globally, kidney diseases were impacting the lives of roughly 850 million people and over 5 million patients died each year due to lack of timely and appropriate treatment [1,2]. Unfortunately, the rampant COVID-19 as a complex multisystem syndrome was implicated in function abnormalities of multiple organs, which would further exacerbate kidney impairment particularly [3,4]. Patients commonly required hemodialysis to eliminate the accumulation of various uremic toxins including water-soluble small-molecular toxins, middle-molecular toxins and protein-bound uremic toxins (PBUTs) from the blood for therapy [5–7]. However, conventional hemodialysis was offered three times weekly and demanded significant volumes of costly dialysate, which heavily limited the miniaturization and portability of hemodialysis device [8,9]. Especially, with the continued global spread of COVID-19, attending to specialized facilities for dialysis treatment not only imposed a burden on the prevention and control of the epidemic, but also increased the infection risk for patients and medical staff rendering the development of home-based dialysis systems become more pressing in quarantine [10,11]. As a result, it was critical to miniaturize the hemodialysis device to enable family dialysis convenient and economical.

With the development of portable artificial kidneys (PAK), different adsorbents layers were placed in the dialysate to adsorb various uremic toxins for continuously keeping the free toxin concentration at low level and dialysis driving force at high altitude, which had become an effective way to substitute vast quantities of dialysate [9,12]. Among the various toxins, PBUTs such as IS and HA were relatively special type of toxins with two different typical forms, of which one was conjugated to important protein and the other was free toxin in blood [13]. There was a dissociative equilibrium between the two forms of toxin suggesting that the equilibrium shifted towards dissociation of the complex prompting the release of more free toxins when part of free toxin was eliminated [13,14]. In this regard, to achieve efficient removal of PBUTs, the concentration of free toxins in the dialysate must be low so that the free fraction in the plasma could diffuse rapidly into the dialysate, which also promoted a shift in equilibrium to dissociation facilitating the continued release of toxin bounded to protein into the plasma.

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and then further penetrated into dialysate [14,15]. Indeed, the introduction of effective adsorbents targeting PBUTs in dialysate not only enhanced the clearance of PBUTs from blood, but also followed and expanded the route of PAK development. Recently, Li et al. proposed that using cyclodextrin-based polymers to adsorb toxins from dialysate could enhance the clearance of PBUTs, but the adsorbent had cumbersome preparation process and tend to dissolve [16,17]. Moreover, in our group, we put forward the concept of lightweight hemodialysis and reduced the usage of the dialysate to one tenth of its original volume by introducing toxin adsorbent with small amount whatever in weight or in volume (almost negligible vs the amount of the dialysate) into the dialysate [9,18]. And we had also attempted to achieve effective clearance of IS by hemodialysis through the design of reduced graphene oxide (rGO) aerogels combined with thin-film nanofibrous composite membrane while the adsorption capacity for PBUTs remained to be improved [13]. Consequently, it was desirable, but challenged to design and develop an easily prepared, highly efficient and low-cost adsorbent for PUBTs.

Concerning potential adsorbent materials or precursors, LDH as a typical anion clay possessed unique structure consisting of positively charged brucite-like layers and intercalated anions [19]. Specifically, the metal layers hold a positive charge due to the fractional isomeric substitution of divalent with trivalent cations while the intercalated properties including simple synthesis, versatile components and structure, great surface adsorption capacity and strong anion exchange performance [25–28]. On the other hand, some researchers had synthesized different LDH with distinct layer spacing and variable morphology by adding different anions to the reaction system, which had contributed to improvements in adsorption performance [26,27]. On the other hand, LDO as a calcined derivative of LDH was also a promising candidate for adsorbent with normally composed of mixed metal oxides or spinel [29]. And the calcination would allow the majority of the water molecules and anions in the LDH layer to be evaporated or decomposed, which could facilitate the enhancement of surface defects and internal reactions [30,31]. Interestingly, LDO exhibited a fascinating property known as the memory effect implying that it tended to reconstruct its original lamellar structure by spontaneous rehydration and simultaneous capture anions into the interlayer in aqueous media [32,33]. Consequently, LDO demonstrated larger specific surface area and superior anion capture performance compared to LDH making it proficiently remove organic and inorganic contaminants in wastewater, such as Cr (VI), U(VI), phosphate, and anionic dyes [19,31,34–36]. However, there were no studies on the application of LDO for the PBUTs adsorption, in spite of the fact that many advantages associated with LDO distinguished it as an idea toxin adsorbent.

Based on the above summary and inspiration, a hierarchically structural flower-like T@MgAl LDH was prepared by a facile one-pot hydrothermal reaction using Mg$^{2+}$ and Al$^{3+}$ as metal sources and citrate as an anionic additive. Evidently, the specific surface area of T@MgAl LDH was much higher than that of conventional bulky MgAl LDH without citrate. And then, the optimized T@MgAl LDH was converted to T@MgAl LDO by calcination maintaining its original morphology entirely, as shown in Scheme 1. Interestingly, T@MgAl LDO was evidenced to possess stretchable nanopores and exhibit great adsorption capacities for PBUTs due to internal intercalation effect and surface adsorption summarized by the calculation of molecular size and inorganic contaminants in wastewater, such as Cr (VI), U(VI), phosphate, and anionic dyes [19,31,34–36]. However, there were no studies on the application of LDO for the PBUTs adsorption, in spite of the fact that many advantages associated with LDO distinguished it as an idea toxin adsorbent.

![Scheme 1. Schematic diagram of the preparation procedure of bulky MgAl LDH, flower-like T@MgAl LDH and T@MgAl LDO.](image-url)

2. Experimental section

2.1. Materials

Trisodium citrate (Na$_3$C$_6$H$_5$O$_7$·2H$_2$O), Al(NO$_3$)$_3$·9H$_2$O, Mg(NO$_3$)$_2$·6H$_2$O, urea, potassium bromide (KBr), sodium oxalate and sodium gluconate were obtained from Sinopharm Chemical Reagent Co., ltd. Indoxyl sulfate potassium salt (IS) and hippuric acid (HA) were purchased by Sigma Aldrich. Deionized (DI) water was used throughout the whole experiments process. All reagents were used as received with no further purification.

2.2. Preparation of LDH and LDO

MgAl LDH was prepared through one-pot hydrothermal reaction
described in the literature with slight modification [37]. Firstly, Al (NO$_3$)$_3$·9H$_2$O, Mg(NO$_3$)$_2$·6H$_2$O and urea were fully dissolved in DI water. And then, different amounts of TSC were added to the solution and stirred thoroughly for 1 h. Finally, the mixture was transferred into a 100 ml autoclave with Teflon-lined and heated at 180°C for 6 h. The resultant products were obtained by centrifugated at 8000 rpm for 5 min when the reaction was completed and cooled. The specific composition ratios and product sample names were shown in Table 1. Moreover, based on subsequent characterization, the T3@MA was chosen to be a paradigm and calcined at 500°C for 3 h in muffle furnace to prepare MgAl-LDO labeled as CT3@MA.

### Table 1

| Sample | Al$^{3+}$ (mmol) | Mg$^{2+}$ (mmol) | urea (mmol) | TSC (mmol) |
|--------|-----------------|-----------------|-------------|-----------|
| MA     | 4               | 4               | 12          | 0         |
| T1@MA  | 4               | 4               | 12          | 0.4       |
| T2@MA  | 4               | 4               | 12          | 0.8       |
| T3@MA  | 4               | 4               | 12          | 1.2       |
| T4@MA  | 4               | 4               | 12          | 1.6       |

#### 2.3. Adsorption experiment

To verify that CT3@MA had a memory effect with stretchable nanoparticles, it was subjected to adsorption experiments for different sized anions. In detail, 40 mg of CT3@MA were dispersed in 0.1 M solutions of KBr, sodium oxalate and sodium gluconate with beaker, respectively. The beaker was fixed on thermostatic shaker at 100 rpm for 24 h. And the temperature was kept at 25°C. After the adsorption was completed, the concentration of the remaining solution was determined with UV spectrophotometer and calculated according to standard curves. Thus, the adsorption capacity would be confirmed from variation in concentration at different time.

In order to explore the adsorption behavior of CT3@MA for PBUTs, IS and HA were used as example toxin adsorbates for the following experiments. To concretely describe the adsorption equilibrium of PBUTs on CT3@MA, 5 mg CT3@MA was put in 50 ml PBUTs aqueous solution in a beaker, respectively. The beaker fixed on thermostatic shaker at 100 rpm for 24 h. And the temperature was kept at 25°C. The initial concentration range for PBUTs was 50–400 mg/L. After adsorption was completed, the concentration of the remaining solution was determined with ultraviolet (UV) spectrophotometer and calculated according to standard curves. The UV absorbance wavelength of IS and HA were 278 nm and 229 nm, respectively. The amount of toxin captured by CT3@MA would be estimated by the difference in concentration of the toxin solution before and after adsorption. And the adsorption capacity was presented in $q_e$ (mg/g) depending on the equation (1):

$$ q_e = (C_0 - C_e) \times V/W $$

where $C_0$ was the initial concentration of PBUTs (mg/L) and $C_e$ was the equilibrium concentration of PBUTs (mg/L). $V$ (L) was the volume of PBUTs solution and $W$ (g) was the mass of CT3@MA.

The Langmuir and Freundlich isotherm model were defined by following equation (2) and (3) [38]:

$$ \frac{C_e}{q_e} = \frac{C_0}{q_m} + \frac{1}{b q_m} $$

$$ \ln q_e = \ln K_F + \frac{1}{n} \ln C_e $$

where $q_e$ (mg/g) and $C_e$ (mg/L) was the adsorption capacity and concentration of PBUTs residual solution at adsorption equilibrium. The $q_m$ (mg/g) was the maximum adsorption capacity. In addition, $b$ (L/mg) expressed the Langmuir constant relevant to adsorption enthalpy. $K_F$ (mg/g) as an empirical constant of Freundlich was adsorption capacity determined by the bond strength and $1/n$ reflected the surface homogeneity and capture strength.

To study the kinetic adsorption of PBUTs on CT3@MA, 50 mg CT3@MA was soaked in 500 ml PBUTs aqueous solution (300 mg/L) in beaker, which was also placed in a thermostatic shaker at 100 rpm for different time (0–300 min). And the temperature was kept at 25°C. Within the time period given, the concentration of the solution was determined with UV spectrophotometer and calculated according to standard curves. Thus, the adsorption capacity would be confirmed from variation in concentration at different time.

The adsorption kinetics could be elaborated using the pseudo-first-order and the pseudo-second-order kinetic models. And the two kinetics were expressed by two equations (4) and (5) [39]:

$$ \ln (q_e - q_t) = \ln (q_e) - k_1 t $$

$$ \frac{t}{q_e} = \frac{1}{k_2 q^2_e} + \frac{t}{q_e} $$

where $q_e$ (mg/g) was the equilibrium adsorption capacity, and $q_t$ (mg/g) was the capacity at the given time $t$ (min). The $k_1$ (h$^{-1}$) and $k_2$ (g/mg·h) were relevant parameters in the models.

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Fig. 1. The FE-SEM images of MA (A), T1@MA (B), T2@MA (C), T3@MA (D), T4@MA (E) and CT3@MA (F).
LDH and LDO samples were obtained by using N2 adsorption–desorption isotherms relying on a specific surface area analyzer (ASAP 2020, Micromeritics Co., USA). The molecular structure and size calculation were provided by Chem 3D combined with Multiwfn where the subfunction of main function 100 was used [40]. The absorbance of the solution of HA or IS was determined by UV–vis spectrophotometer (TU-1950, Perkinje, China).

3. Results and discussion

3.1. Characterisation of LDH and LDO

3.1.1. Morphology and composition characterisation

Under high temperature conditions, Mg2+ and Al3+ were formed into MgAl LDH by hydrothermal reaction. The introduction of citrate into the system would lead to interesting changes in the morphology of the product after reaction, and the effect of adding amounts of TSC on the morphology was investigated in detail. As could be seen in Fig. 1A-B, the resultant LDH presented a typical large thick micro-sheet with uneven size without citrate, which would transform into a curled and three-dimensional lamellar structure in the presence of citrate. Moreover, as the amount of citrate increased, the nanoplate became smaller and more well-defined, eventually emerging in stacks (Fig. 1C) and assuming a complete flower-like spherical shape with delicate hierarchically structure (Fig. 1D-E). As T4@MA and T3@MA were extremely similar in morphology, T3@MA was chosen for the subsequent study. Thus, after calcination at 500°C, it was converted to LDO and the morphology remained as a flower-like sphere with a slight increase in dispersion, as shown in Fig. 1F.

To confirm the introduction and removal of citrate, the FT-IR spectrum of MA, T3@MA and CT3@MA was demonstrated in Fig. 2. The characteristic peak at 3448 cm⁻¹ was the stretching vibration of –OH attributed to interlayer water molecules and LDH layer (H-OH and Mg/Al-OH) [41,42]. In particular, the characteristic peak at 1356 cm⁻¹ was ascribed to the C–O groups of CO₃²⁻ in interlayer [31]. And the bands observed from 500 to 800 cm⁻¹ were assignable to the vibrations of Mg-O/Al-O [43]. For T3@MA, the obvious two peaks at 1619 cm⁻¹ and 1401 cm⁻¹ belonged to the asymmetric and symmetric stretching vibrations of COO⁻ related to citrate, respectively [44,45]. In addition, the intensity of peaks assigned to –OH and COO⁻ were both significantly decreased suggesting that the interlayer citrate was removed by decomposition during calcination.

3.1.2. Shaping principle and stretchable nanopores

In order to investigate the intercalation effect of citrate, XRD patterns of a series of products were shown in Fig. 3A. It was clearly that the sharp reflection peaks of MA were appeared at 11.73°, 23.52°, 34.94°, 39.52° and 47.03°, corresponding to the crystal planes of (0 0 3), (0 0 6), (0 1 2), (0 1 5) and (0 1 8), respectively [46,47]. Obviously, all the characteristic peaks relating to LDH reappeared for T3@MA, but the peak positions had migrated towards lower angles [48], where the reflection peaks related to (0 0 3) and (0 0 6) crystal planes transformed to 7.35° and 14.67°. Hereby, it implied a broadening of basal spacing, suggesting the intercalation of citrate into the layers of T3@MA. Specifically, the basal spacing of the (0 0 3) reflection of MA and T3@MA was 7.54 Å and 12.00 Å by calculation. In addition, the thickness of LDH metal layer was already known to be 4.8 Å [25,49]. The space occupied by citrate could be easily deduced to be 7.2 Å, which was in perfect agreement with the molecular size of citrate obtained from Chem 3D combined with Multiwfn as displayed in Fig. S1A-B. Therefore, it was implying that the arrangement of citrate in the LDH interlayer might be a monolayer state and perpendicular to the host layer shown in Fig. S1C.

Fig 2. The FT-IR spectrum of MA, T3@MA and CT3@MA.

Fig 3. The XRD patterns of MA, T1@MA, T2@MA, T3@MA, T4@MA and CT3@MA.
characteristic peaks of LDH all disappeared after calcination while the late (Ox-CT3@MA) and gluconate (Gl-CT3@MA).

Fig. 4. -

S. Ding et al.

The XRD patterns of CT3@MA adsorbed bromide (Br-CT3@MA), oxalate (Ox-CT3@MA) and gluconate (Gl-CT3@MA).

Furthermore, XRD patterns of LDH synthesized in Fig. 3B by different TSC amount were also characterized for exploiting the role of citrate. Evidently, the peak attributed to the (0 0 3) crystal plane of T1@MA was at 11.71°, which was in accordance with that of MA. It was predicting that the citrate had not entered the interlayer, but the morphology of T1@MA was dramatically different from that of MA according to Fig. 1A-B. Nevertheless, the peak attributed to the (0 0 3) crystal plane of MA, T3@MA and CT3@MA were 11.10°, 9.15° and 7.35° respectively. And the basal spacing of the (0 0 3) reflection of them were 7.96 Å, 9.64 Å and 12.01 Å by calculation. Thus, the space occupied by bromide, oxalate and gluconate also could be deduced to be 3.16 Å, 4.84 Å and 7.21 Å respectively, which were all slightly smaller than their molecular size which could be attributed to the insertion into the layers at an angle to the crystal plane. Therefore, the insertion of various anions allowed for different crystal spacing indicating that CT3@MA had adaptable and stretchable nanopores.

3.1.3. N_2 adsorption–desorption

N_2 adsorption–desorption isotherms would reveal the porosity of the product rely on the obtained specific surface area and pore size, as shown in Fig. 5. Theoretically, the specific surface area of the hierarchical flower-like spheres was expected to be much higher than that of the micro-sheet alone, and would increase further after calcination. In Fig. 5A, the isotherms of MA, T3@MA and CT3@MA were categorized as type IV isotherm indicating that the sample contained mesopores [57]. In addition, the hysteresis loop as type H3 was obviously shown in the isotherms of T3@MA and CT3@MA owing to the aggregation and stack of nanoplates [35]. From Fig. 5B, it could be concluded that T3@MA had more pores concentrated below 10 nm compared to MA, but also possessed larger pores (20–40 nm) due to the stacking of Mg²⁺ and Al³⁺ resulting in a lower rate of nucleation and crystallization of the LDH and then diminished the size and thickness of the product. Also, the surface energy of individual nanosheets was high from thermodynamic perspective and the optimum approach to reducing the surface energy of nanosheets was effective and ordered agglomeration [51,52]. Thus, the nanosheets became smaller as the citrate addition increased, and eventually the small-sized irregular flakes spontaneously self-assembled into stacks of completed spherical particles. In summary, citrate was inserted between the layers of LDH as a complexing agent and acted as a linker to facilitate the assembly of the flakes during the reaction. Analogously, citric had been employed extensively as a crystallization control or bridging agent in the synthesis of nanoscale metal compounds in a broad variety of shapes and structure [53–56].
nanoplates. Moreover, CT3@MA had an increased pore size and a wider distribution. And the detailed data were listed in Table 2. The specific surface area of T3@MA was four times larger than that of MA and hold an enhanced average pore size and total pore volume. Furthermore, the specific surface area, average pore size and total pore volume of CT3@MA was also enlarged due to the removal of the majority of interlayer water and citrate by calcination. Therefore, CT3@MA was expected to hold an abundance of adsorption sites for toxins.

3.2. Adsorption isotherms and kinetics

LDO had a higher specific surface area and anion complexation capacity compared with LDH, so that CT3@MA would be capable of capturing HA or IS. For the purpose of predicting the maximum adsorption capacity and investigating the potential interaction between PBUTs and CT3@MA, the adsorption isotherm models such as Langmuir and Freundlich were employed to analyze the adsorption experiment data based on the relationship between the adsorption capacity and the equilibrium concentration of adsorbate. And the appropriate model was determined by the correlation coefficient (R²) of the fitting equation. For the Langmuir model, there was an assumption that only a limited number of adsorption sites could capture the adsorbate. And the adsorption process only occurred as mono-layer, so that only one molecule of adsorbate was attached to each active site. Hence, once all the adsorption sites were occupied, the adsorbent would be unable to take up the adsorbate resulting in a saturated adsorption capacity. Conversely, the Freundlich model posited that as the concentration of adsorbate increased, the adsorbent surface allowed the adsorbate to be trapped by multi-layer adsorption, which was widely used to account for heterogeneous systems.

The effect of initial PBUTs concentration on the equilibrium adsorption capacity was shown in Fig. 6. Normally, the uptake of both HA or IS by CT3@MA increased significantly with increasing initial concentration of them. After a certain concentration was reached, the curve was often gently flattened to a plateau. From Fig. 6A, the equilibrium adsorption capacity of CT3@MA for HA and IS was 119.8 mg/g

| Sample   | S_{BET} (m²/g) | V_{tot} (cm³/g) | D_{ave} (nm) |
|----------|----------------|-----------------|--------------|
| MA       | 37.7           | 0.260           | 3.408        |
| T3@MA    | 158.6          | 0.613           | 3.827        |
| CT3@MA   | 187.3          | 1.011           | 5.642        |

Table 3

| Adsorbates | Langmuir constants | Freundlich constants |
|------------|--------------------|----------------------|
|            | q₀ (mg/g) | B (L/mg) | R² | kₐ (mg/g) | n | R² |
| IS         | 63.1      | 0.031  | 0.9899 | 9.3     | 3.01 | 0.7575 |
| HA         | 129.8     | 0.048  | 0.9941 | 21.3    | 3.00 | 0.8701 |

Fig. 6. (A) Effect of initial concentration on the adsorption capacity of the CT3@MA for IS and HA. Adsorption isotherms of IS and HA depending on the Langmuir equation (B) and the Freundlich equation (C).
and 57.5 mg/g, corresponding to equilibrium concentration of 250 mg/L and 200 mg/L, respectively. As illustrated in Fig. 6B-C, the adsorption curves were more closely aligned to the Langmuir fitting trend than that to Freundlich for both HA and IS. The parameters of the adsorption kinetics were presented in Table 3. Clearly, the adsorption process of HA or IS on CT3@MA was better described by the Langmuir model, indicating that the adsorption process was a mono-layer adsorption procedure on the homogeneous adsorbent. And the maximum adsorption capacity of HA and IS was up to 129.8 mg/g and 63.1 mg/g, respectively. Compared to conventional carbon materials and some adsorbents used in dialysate shown in Table 4, such impressive adsorption capacity indicated that CT3@MA would be utilized effectively for removing PBUTs.

To evaluate the adsorption rates of HA or IS on CT3@MA, the toxin adsorption experiments were carried out over a time range of 10 min to 300 min, both at an initial concentration of 300 mg/L. As shown in Fig. 7A, the adsorption capacity increased rapidly within the first 1 h, at which moment approximately 90 % of the adsorption capacity was accomplished because of the abundance of vacant adsorption sites in the initial phase. After that, the adsorption process continued in a slow growth period due to the reduction of the available adsorption sites until the plateau was established, which represented the completion of the adsorption. And the adsorption procedure reached equilibrium within 4 h, which exactly matched the treatment duration of clinical hemodialysis. In addition, the pseudo-first-order kinetic model and pseudo-second-order kinetic model were deployed to fitting experimental data to investigate further the adsorption process. Likewise, the best applicable model was determined by the correlation coefficient ($R^2$) of the two models. The kinetic fitting plots were depicted in Fig. 7B-C and the parameters were available in Table 5. It was clear that the $R^2$ of the curves fitted with the pseudo-second-order kinetic model were much higher than those with the pseudo-first-order kinetic model.

### Table 4

| Adsorbents                        | Capacity (mg/g) |
|----------------------------------|-----------------|
| CMK-3 (mesoporous carbon) [58]   | 12.2            |
| Norit (activated charcoal) [58]  | 12.6            |
| PES-zeolite composite membrane   | \               |
| rGO aerogel [13]                 | \               |
| poly-cyclodextrins [16]          | \               |
| This work CT3@MA                 | 129.8           |

### Table 5

| Adsorbates | Pseudo-first-order model | Pseudo-second-order model |
|------------|--------------------------|---------------------------|
|            | $q_e$ (mg/g) | $K_1$ (1/h) | $R^2$ | $q_e$ (mg/g) | $K_2$ ((g/(mg h)) | $R^2$ |
| IS         | 26.2         | 0.0096     | 0.8423 | 58.1         | 0.0043           | 0.9991 |
| HA         | 19.4         | 0.0170     | 0.4819 | 123.4        | 0.0001           | 0.9989 |

Fig. 7. (A) Effect of contact time on the adsorption capacity of the capacity of the CT3@MA for IS and HA. Pseudo-second-order kinetic model (B) and Pseudo-first-order kinetic model (C) for the adsorption process.
higher than that fitted with the pseudo-first-order kinetic model for HA and IS. As expected, the fitted equilibrium adsorption capacity based on pseudo-second-order kinetic model (123.4 mg/g and 58.1 mg/g) was much closer to real experimental data (119.8 mg/g and 57.5 mg/g) than that on pseudo-first-order kinetic mode (19.4 mg/g and 26.2 mg/g), which implied that the adsorption of HA and IS was a classical chemical.

Fig. 8. The SEM image of IS-CT3@MA (A) and HA-CT3@MA (B).

Fig. 9. The Mg 1s (A) and Al 2p (B) spectrum of CT3@MA, IS-CT3@MA and HA-CT3@MA.
interaction process. And the adsorption mechanism would be discussed in detail later.

According to our previous work [9,18], we had successfully reduced the volume of the dialysate to one tenth of its original volume by introducing toxin adsorbent into the dialysate. Obviously, small mass adsorbent (2 g) was able to replace large volume dialysate (1800 ml) while the toxin (creatinine) clearance remained at a high level (63 %), which illustrated the feasibility of lightweight hemodialysis and the importance of the adsorbent in the whole process [18]. For PBUTs, we had similarly performed rGO aerogel blocks as adsorbents into the dialysate to achieve efficient removal of IS with limited amounts of dialysate (200 ml) during the hemodialysis, of which the toxin clearance was increased from 36.5 % to 51.8 % with the assistance of adsorbents [13]. In this study, CT3@MA not only exhibited the maximum adsorption capacity (63.1 mg/g) for IS comparable to that of rGO aerogel (69.4 mg/g) and reached adsorption equilibrium (3 h) faster than rGO aerogel (4 h) [13], but also possessed a high adsorption capacity for HA (129.8 mg/g) and the adsorption equilibrium time was only 2 h. Therefore, the CT3@MA as an adsorbent towards PBUTs provided great adsorption capacity and fast adsorption rate, which would definitely perform more effective removal of toxins in the same way indicating that it had great potential to be employed in family hemodialysis equipment.

3.3. Adsorption mechanism

In attempt to elucidate the mechanism of adsorption of PBUTs in detail, CT3@MA was characterized and analyzed by FE-SEM, XPS and XRD before and after adsorption. And samples after adsorption of HA and IS were labelled as HA-CT3@MA and IS-CT3@MA, respectively. On the one hand, the large specific surface area of CT3@MA inevitably provided abundant active sites for toxin adsorption. On the other hand, the extensive pore structure formed by the stacking of nanoplates also had a certain physical binding effect on toxins. In Fig. 8A-B, it was clear that the flower-like morphology of the nanoplates stacks remained after the adsorption of the PBUTs, proving the good stability of the LDO. However, the surface of the nanoplate showed abundant tiny folds, which may be the morphology formed by the micro-crystallization of HA or IS.

The XPS survey spectrums of samples were shown in Fig. S3 and Fig. 9. As seen from Fig. S3A, in contrast to CT3@MA, a few new elements of N and S appeared after adsorption. In addition, As shown in Fig. S3B-D, The peak with the binding energy 399.5 eV was attributed to N—C while the peaks with the binding energy 168.3 eV and 169.3 eV were attributed to S 2p3/2 and S 2p1/2 respectively from -SO3-[60]. And in Fig. 9, it could be seen that the peak at 1304.0 eV and 74.1 eV were assigned to Mg-O and Al-O consistent with the formation of Mg and Al bimetallic oxides [61,62], which also demonstrate MgAl LDO was successfully prepared. Moreover, the peaks of Mg-O and Al-O in IS-CT3@MA and HA-CT3@MA were slightly shifted to the high binding energy indicating that there was electrostatic interaction of the PBUTs with the bimetallic oxides due to anisotropic charges. This provided ample evidence for the adsorption of PBUTs on CT3@MA.

Moreover, LDO was capable of incorporating anions into their layers to re-generate LDH structures through the so-called memory effect with stretchable nanopores. As shown in Fig. 10, it was found that the layered structure of the LDH material was reconstructed, which could be deduced from the proximity of its XRD pattern to MA and CT3@MA. Interestingly, the peak attributed to (003) plane was at 7.33° and 7.17°
for HA-CT3@MA and IS-CT3@MA respectively, which demonstrated that the crystal plane spacing could be calculated to be 12.04 Å and 12.37 Å. Likewise, the space occupied by HA and IS could be easily deduced to be about 7.24 Å and 7.57 Å, respectively. Furthermore, the molecule size of HA and IS was shown in Fig. 11. By Chem 3D combined method followed by calcination treatment, providing a wealth of active sites for adsorption due to the ultrahigh specific surface area. Moreover, taking advantage of the unique memory effect and stretchable nanomaterials, the maximum adsorption capacity of HA and IS was 129.8 mg/g and 63.1 mg/g respectively relying on electrostatic interaction to capture toxins by surface adsorption and interlayer intercalation. In summary, the design and application of such LDO not only provided a new reference for the construction of nano-materials with hierarchical structure and stretchable nanopores using citrate as the key role of the surfactant template, Langmuir 35 (2019) 13562–13569.

4. Conclusion

In this work, we had prepared flower-like spherical MgAl LDO with delicate hierarchical structure and stretchable nanomaterials for dialysis/adsorption bifunctional thin-film nanocomposite membrane for creatinine clearance in portable artificial kidney. The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

Data availability

Data will be made available on request.

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Appendix A. Supplementary material

Supplementary data to this article can be found online at https://doi.org/10.1016/j.seppur.2022.122033.

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