INTRODUCTION

Parasitism is a potent source of selection in natural host populations and has recently been suggested to play a role in processes such as maintenance of sexual reproduction (Jokela, Dybdahl, & Lively, 2009; King, Delph, Jokela, & Lively, 2009) and divergence of host populations (Eizaguirre, Lenz, Kalbe, & Milinski, 2012; Karvonen & Seehausen, 2012). Indeed, several parasite taxa impair host condition through depletion of resources, tissue damage, and manipulation of host behavior (Barber, Hoare, & Krause, 2000; Barber & Svensson, 2003; Hafer & Milinski, 2016; Jokela, Taskinen, Mutikainen, & Kopp, 2005; Karvonen, Seppälä, & Valtonen 2004a; Moore, 2002; Seppälä, Liljeroos, Karvonen, & Jokela, 2008) and thus have severe implications for host fitness. The risk of parasitism is often structured both spatially and temporally because of spatial aggregation of infected individuals and parasite intermediate hosts (Byers, Blakeslee, Linder, Cooper, & Maguire, 2008; Faltýnková, Valtonen, & Karvonen, 2008; Jokela & Lively, 1995; Karvonen, Cheng, & Valtonen, 2005), and
Several host species of parasites express secondary sexual characteristics through which they can advertise their vigour, as well as resistance to parasites (Hamilton & Zuk, 1982). Often these signals are visually perceived ornaments such as long tails or bright coloration, and commonly displayed by males. For example, the connection between the expression of sexual ornaments and parasitism has been demonstrated in many species of birds (Hörak, Ots, Vellau, Spottiswoode, & Pape Møller, 2001; McGraw & Hill, 2000; Thompson, Hillgarth, Leu, & McClure, 1997) and fish (Barber, Arnott, Braithwaite, Andrew, & Huntingford, 2001; Houde & Torio, 1992; Maan, van der Spoel, Jimenez, van Alphen, & Seehausen, 2006).

Overall, current evidence strongly suggests that male sexual ornaments could signal resistance to parasitism. However, reduction in host vision could impair the ability of individuals to perceive sexual signals. For example, it has been shown in cichlid fishes that visually perceived sexual signals advertised through male coloration can be blurred because of increased water turbidity, resulting in hybridization of species following the relaxation of color-based sexual selection (Seehausen, van Alphen, & Witte, 1997). Similarly, parasites found in the key sensory organs, such as the eyes, could impair host vision and the ability to perceive visual cues from potential mates (Karvonen & Seehausen, 2012). In species where sexual selection is based on males displaying secondary sexual characteristics, the ability of females to judge male quality when infected with such parasites could be compromised. However, while mating decisions are known to be influenced by general condition (Cotton, Small, & Pomiankowski, 2006) and also parasite infections (Lopez, 1999; Mazzi, 2004; Pfennig & Tinsley, 2002; Poulin & Vickery, 1996) of the choosier sex, the potential of parasites directly interfering with sex-related traits and possible gender differences in parasitism. We were particularly interested in variation in infection of parasites inhabiting sensory organs of fish that could show the potential of parasite-induced changes in conditions of sexual selection and mate choice among the sampling locations. Similarly, infections could influence overwinter survival of the fish and result in lower infection levels in the early summer. Furthermore, to explain patterns of infection of the parasites in the wild, we exposed male and female sand gobies to controlled experimental infection from trematode eye flukes in the laboratory. Differences in the abundance of these parasites between male and female fish under similar level of exposure would be consistent with the idea of differences in susceptibility between the sexes.

2 | MATERIALS AND METHODS

2.1 | Sampling of gobies

Sand gobies and common gobies were sampled from three locations in the proximity of the Tvärminne Zoological station, southern Finland. The first location was next to the station (referred here to as “Station”; 59°50'41"N, 23°14'58"E), the second ca. 500 m from the Station (Långholmen; 59°50'48"N, 23°15'12"E), and the third farther out to the sea ca. 4 km from the Station (Vargskär; 59°49'24"N, 23°08'38"E). All locations had a bottom substrate of sand or soft mud with little or no vegetation. Sampling was conducted three times, June 2014, June 2015, and October 2015, to capture both spatial and temporal variation in infections. At each location, gobies were caught from a depth of 0.5–1 m using a seine net. However, sampling campaigns at some of the locations were unsuccessful at times (Table 1) due to complete absence of fish. Fish were brought alive to the laboratory, euthanized, sexed, measured for length (mm), and inspected fresh for infections on fins (right pectoral fin), gills, eyes, and internal organs under a microscope. Parasites were identified at genus or species level when possible. Prevalence (proportion of fish infected, %) and mean abundance (mean parasite...
**TABLE 1** Prevalence (% fish infected) and mean abundance (number of parasites per fish ± SE) of the parasite taxa detected in sand gobies captured from three locations in the Northern Baltic Sea in June 2014–October 2015

| Parasite            | June 2014 |          | June 2015 |          | October 2015 |          |
|---------------------|-----------|----------|-----------|----------|--------------|----------|
|                     | Station   | Långholmen | Vargskär  | Station   | Långholmen | Vargskär  | Station   | Långholmen | Vargskär  |
|                     | n = 29    | n = 30   | n = 31    | n = 30   | n = 30     | n = 30    | n = 0     | n = 6      | n = 14    |
| Trichodina sp.      | 90        | 86.7     | 67.7      | 96.7     | 96.7       | 66.7      | 16.7      | 57.1       | 0.2 ± 0.2  |
|                     | 11.2 ± 3.4| 2.9 ± 0.4| 2.6 ± 0.9 | 24.6 ± 6.6| 14.0 ± 2.8 | 5.6 ± 1.8 | 0.2 ± 0.2 | 1.7 ± 0.7  |
| Gyrodactylus sp.    | 0.0       | 10       | 0.1 ± 0.1 | 10       | 3.3        | 10        | 0.0       | 57.1       | 0.9 ± 0.4  |
|                     | 0.1 ± 0.1 | 0.0      | 0.03 ± 0.03| 0.4 ± 0.1 | 0.2 ± 0.1  | 0.4 ± 0.1 | 0.2 ± 0.2 | 2.8 ± 11.3 |
| Diplostomum spp.    | 0.0       | 6.7      | 3.2       | 26.7     | 23.3       | 30        | 16.7      | 78.6       | 0.9 ± 0.4  |
|                     | 0.07 ± 0.05| 0.03 ± 0.03 | 0.4 ± 0.1 | 0.2 ± 0.1 | 0.4 ± 0.1  | 0.2 ± 0.2 | 2.8 ± 11.3 |
| Diplostomum baeri  | 3.4       | 6.7      | 3.3       | 6.7      | 3.3        | 0.0       | 0.0       | 21.4       | 0.4 ± 0.2  |
|                     | 0.07 ± 0.07| 0.03 ± 0.03 | 0.07 ± 0.05| 0.03 ± 0.03 | 0.03 ± 0.03 | 0.03 ± 0.03 | 0.03 ± 0.03 | 0.03 ± 0.03 |
| Apatemon sp.        | 31        | 33.3     | 9.7       | 40       | 30         | 40        | 16.7      | 64.3       | 0.2 ± 0.2  |
|                     | 0.4 ± 0.1 | 0.8 ± 0.3 | 0.1 ± 0.1 | 1.0 ± 0.3 | 0.5 ± 0.2  | 0.9 ± 0.3 | 0.2 ± 0.2 | 2.9 ± 1.6  |
| Cryptocotyle sp.    | 56        | 80       | 80.6      | 50       | 26.7       | 70        | 0.0       | 71.4       | 2.2 ± 0.5  |
|                     | 0.9 ± 0.2 | 2.9 ± 0.5 | 5.6 ± 1.0 | 1.2 ± 0.3 | 0.5 ± 0.2  | 6.2 ± 1.6 | 0.0       | 2.2 ± 0.5  |
| Sphaerostomum sp.   | 0.0       | 6.7      | 0.0       | 0.0      | 0.0        | 0.0       | 0.0       | 0.0        | 0.0       |
|                     | 0.07 ± 0.05| 0.03 ± 0.03 | 0.07 ± 0.05| 0.03 ± 0.03 | 0.03 ± 0.03 | 0.03 ± 0.03 | 0.03 ± 0.03 | 0.03 ± 0.03 |
| Proteocephalus sp.  | 13.8      | 13.3     | 6.5       | 3.3      | 3.3        | 3.3       | 0.0       | 0.0        | 0.0       |
|                     | 0.2 ± 0.1 | 0.1 ± 0.1 | 0.1 ± 0.1 | 0.03 ± 0.03| 0.03 ± 0.03 | 0.03 ± 0.03 | 0.03 ± 0.03 | 0.03 ± 0.03 |
| Camallanus lacustris| 3.4       | 3.3      | 0.0       | 0.0      | 0.0        | 0.0       | 0.0       | 0.0        | 0.0       |
|                     | 0.03 ± 0.03| 0.07 ± 0.07| 0.07 ± 0.07| 0.03 ± 0.03| 0.03 ± 0.03 | 0.03 ± 0.03 | 0.03 ± 0.03 | 0.03 ± 0.03 |
| Unidentified nematode| 10.3     | 16.7     | 16.1      | 6.7      | 0.0        | 6.7       | 0.0       | 7.1        | 0.07 ± 0.07 |
|                     | 0.1 ± 0.1 | 0.2 ± 0.1 | 0.2 ± 0.1 | 0.07 ± 0.05| 0.07 ± 0.05| 0.07 ± 0.05| 0.07 ± 0.05| 0.07 ± 0.07|
| Echinorhynchus gadii| 6.9       | 0.0      | 0.0       | 0.0      | 0.0        | 0.0       | 0.0       | 0.0        | 0.0       |
|                     | 0.1 ± 0.1 | 0.1 ± 0.1| 0.1 ± 0.1| 0.1 ± 0.1| 0.1 ± 0.1  | 0.1 ± 0.1| 0.1 ± 0.1| 0.1 ± 0.1  |
| Neoechinorhynchus rutili| 6.9     | 0.0      | 0.0       | 0.0      | 0.0        | 0.0       | 0.0       | 0.0        | 0.0       |
|                     | 0.07 ± 0.05| 0.2 ± 0.1 | 0.2 ± 0.1| 0.1 ± 0.1| 0.1 ± 0.1  | 0.1 ± 0.1| 0.1 ± 0.1| 0.1 ± 0.1  |
number per fish) were calculated for each parasite taxa. In addition, fish eye lenses were studied before dissection for coverage of cataracts caused by *Diplostomum* spp. eye flukes using slit-lamp (Kowa SL-15) microscopy (Karvonen et al., 2004a). Cataracts were scored as 10%, 20%, ..., 100% coverage of the lens area, which correlates with the deleterious effects of the parasites on fish (Karvonen & Seppälä, 2008; Seppälä, Karvonen, & Valtosen, 2005) and thus provides an indirect measure of parasite-induced effects on the host. Differences in total parasite abundance between the locations and sampling times were analyzed using GLMs with negative binomial distribution and log link function. Abundances of the parasite taxa *Trichodina* sp. and *Gyrodactylus* sp. (see Results) were excluded from the analyses as they were studied only from one pectoral fin.

### 2.2 | Experimental exposure

Experimental exposure of sand gobies (Figure 1) was conducted in three containers each with 6 l of water (16°C) taken from the Baltic Sea, continuous aeration, and 10 fish (five females and five males) captured 1 week earlier from the Vargskär sampling location, totaling 30 fish (mean length: 48.0 ± 2.0 mm [females], 53.5 ± 1.9 mm [males]). Before the experiment, the fish had been housed in replicated large stock aquaria supplied with a continuous flow of fresh seawater and fed with live mysid shrimp and frozen Chironomidae larvae ad libitum. In addition to the three exposure containers, one container with 10 fish and an even sex ratio served as the unexposed control. This was used to record possible infections resulting from parasite infective stages present in the water, if any, and those that had taken place recently in the wild before the fish were caught, which could not be separated from those resulting from the experimental exposure (see below). Each of the three infection containers then received a total dose of 750 *D. pseudospathaceum* cercariae (75 cercariae per fish) that had been released by five infected *Lymnaea stagnalis* snails collected from Lake Vuojärvi, Central Finland. Note that there is no detectable population genetic structure in these parasites across Finland (Louhi, Karvonen, Rellstab, & Jokela, 2010), which is why parasite origin was unlikely to affect the results. The snails were allowed to produce cercariae for 2 hr in 2 dl of water (20°C). Suspensions of the snails were then combined, and the cercarial density was estimated by taking ten 1 ml samples. Water in the containers was regularly mixed during the first hour of exposure to ensure equal exposure of all individuals to the parasite. The fish were maintained in these conditions for 18 hr, which is sufficient time for the parasites to reach the eye lenses in a small fish (Louhi, Sundberg, Jokela, & Karvonen, 2015). There was no mortality of fish during or after the exposure. All fish were then euthanized and studied for the number of parasites in the eye lenses. Older infections originating from the wild and those originating from the experimental exposure could be distinguished based on the size of the parasite metacercariae. All experimental procedures were in accordance with the ethical standards of the Finnish Regional State Administrative Agency and conducted under license (License code: ESAVI/4706/04.10.07/2015). Data were analyzed using ANCOVA with fish sex as a fixed factor and container as a random factor to account for dependency among fish exposed in the same container. Fish length was used as a covariate. All analyses were conducted using SPSS 24 statistical package.

### 3 | RESULTS

In total, 12 parasite taxa were detected among the 200 sand gobies and 168 common gobies examined, including five nematodes, two acanthocephalans, one cestode, one monogenean, and one protozoan (Table 1, Table 2). All parasite taxa were observed in both fish species, except for the acanthocephalan *Echinorhynchus gadi*, which was observed only in the sand goby. The mean number of parasite taxa per fish was significantly higher in the sand goby (2.23 ± 0.1, all numbers indicate mean ± SE, range 0–7) than in the common goby (1.78 ± 0.1, range 0–4) (t test: *t* = 3.76, *p* < .001; locations and sampling times combined). In both fish species, the protozoan *Trichodina* sp. was the most prevalent and abundant parasite taxa with prevalence ranging between 16.7% and 96.7% in the sand goby and 0%–100% in the common goby, depending on the location and sampling time (Table 1, Table 2). Among the macroparastites, trematodes *Cryptocotyle* sp., *Apatemon* sp., and *Diplostomum* spp. were most commonly observed. Mean total parasite abundance was significantly higher in the common goby compared to the sand goby (GLM: Wald = 11.76, *p* < .001; sampling locations combined) while this depended on the sampling time (Wald = 30.65, *p* < .001 (sampling time); Wald = 32.86, *p* < .001 (species × sampling time); Figure 2). There were also significant differences in parasite abundance among the locations so that Långholmen and Vargskär locations had the highest parasite abundances (GLM: Wald = 14.25, *p* = .001; sampling times combined) while this again depended on the fish species (Wald = 28.01, *p* < .001 (species); Wald = 32.63, *p* < .001 (species × sampling location); Figure 2).

Total parasite abundance was not different between males and females in either fish species (Wald = 1.05, *p* = .305 (sex); Wald = 1.98, *p* = .159 (species × sex); immature fish excluded, sampling locations

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**FIGURE 1** Female sand gobies (*Pomatoschistus minutus*). Photograph by Kai Lindström
and sampling times combined). However, a more detailed analysis of the infections of Diplostomum eye fluke in Vargskär showed that females harbored significantly higher abundances of these parasites compared to males (GLM: Wald = 4.94, $p = .026$ (sex); Wald = 30.64, $p < .001$ (sampling time); Figure 3a). Infections of eye flukes also caused significant cataract coverage in the eye lenses of sand gobies. Cataract coverage increased with parasite abundance so that the coverage rapidly approached 100% when there was more than one parasite per lens. The relationship was captured by a steep asymptotic curve (Figure 4). However, there was no difference in cataracts caused by a given parasite abundance between the male and female sand gobies ($t_{36} = 0.228$, $p = .821$). While this suggests similar susceptibility to parasite-inflicted damage between the sexes, eyes of female sand gobies in Vargskär nevertheless showed higher average cataracts than males because of their higher parasite abundances ($t_{82} = 2.082$, $p = .040$). No cataracts were observed in uninfected eye lenses.

Experimental exposure of sand gobies from Vargskär to *D. pseudospathaceum* infection indicated that all fish became readily infected with the parasite. The mean parasite abundance per fish increased with body size (ANOVA: $F_{1,23} = 15.60$, $p < .001$) and ranged from 43.0 ± 6.0 to 52.6 ± 5.8 depending on the container ($F_{2,21} = 4.06$,
Spatial and temporal variation in parasitism can have significant implications for host populations by underlying variability in influence of parasites on the condition and reproduction of individuals (Lefevre et al., 2009; Tompkins, Dunn, Smith, & Telfer, 2011; Wood et al., 2007). This is particularly clear with infections that can cause detrimental effects on the hosts, such as those located in the central sensory organs. Moreover, hosts species in which such effects can come about already at low parasite numbers may suffer the most notable consequences. We examined parasite infections of two species of gobid fishes in the Northern Baltic Sea. We scored the entire macroparasite community in these fishes and specifically targeted eye flukes of the genus Diplostomum that are known to cause loss of vision in several fish species, with the effects likely being most prominent in smaller fish species with the smallest eye lenses. We found significant spatiotemporal variation in infections as well as differences between the fish species so that the common gobies were more heavily infected in most cases. While there was no effect of host sex on the infections overall, female sand gobies were more heavily infected with eye flukes at the sampling location with the highest infection. An experimental exposure of the fish indicated that the difference between the sexes was at least partly explained by the higher susceptibility of females to infection.

Spatiotemporal variation in parasitism is a common feature of most host–parasite interactions, including parasitic infections of fish in freshwater (Marcogliese, Gendron, Plante, Fournier, & Cyr, 2006; de Roij & MacColl, 2012) and in sea (Grutter, 1998; Sikkel, Nemeth, McCammon, & Williams, 2009). We found that parasite infections of the two goby species followed similar patterns. First, infection abundances were clearly different between the locations, particularly in the common goby, and mainly driven by the trematode Cryptocotyle sp. This could reflect, for example, spatial variation in infection prevalence in the first intermediate snail hosts of the parasite, which is commonly observed also in other trematode systems (Faltýnková et al., 2008; Jokela & Lively, 1995). Second, with few exceptions, directly transmitted parasites (Trichodina sp. and Gyrodactylus sp.) were clearly more prevalent in the early summer compared to autumn, reflecting the temperature-driven replication of these parasites (Bagge & Valtonen, 1999; Halmetoja, Valtonen, & Taskinen, 1992; Koskivaara, Valtonen, & Prost, 1991; Rintamäki-Kinnunen & Valtonen, 1997). Third, trophically transmitted parasites (cestodes, nematodes, and acanthocephalans) showed sporadic occurrence at low numbers, suggesting that the gobies unlikely are the primary fish hosts for these parasites in this system. Fourth, many of the larval trematodes, particularly in sand gobies from the Vargskär location with the highest infection, tended to be more abundant in autumn compared to early summer (Table 1). This well reflects the typical accumulation of trematodes in their intermediate hosts during summer months (Faltýnková, Karvonen, & Valtonen, 2011; Karvonen, Hudson, Seppälä, & Valtonen, 2004; Karvonen et al., 2004a,b). The lower abundance in the early summer is also consistent with the idea that the most heavily infected individuals may be lost from the population during winter, while our data were too few to test this properly. For example, Diplostomum infections and cataracts comparable to this study in the eye lenses are known to cause serious fitness consequences in fish (Crowen & Broom, 1980; Karvonen & Seppälä, 2008; Seppälä et al., 2005), supporting a possibility of parasite-driven population effects (Marcogliese, Compagna, Bergeron, & McLaughlin, 2001). Overall, broad lines of the spatial and temporal variation of parasitism in this system are in accordance with earlier findings of parasite infections in gobies in the Baltic Sea (Zander, 2003, 2005; Zander & Kesting, 1998; Zander, Strohbach, & Gorenwold, 1993).

We also found a distinct difference between male and female sand gobies in the abundance of eye fluke infection so that females were more heavily infected both in the field and in the experimental exposure. This is in contrast with the general pattern of higher infestation in males across a range of host–parasite systems (Klein, 2004; Poulin, 1996), although few studies have reported higher parasite infections also in females, for example, in guppies (Richards, van Oosterhout, & Cable, 2010; Stephenson, van Oosterhout, Mohammed, & Cable, 2015; Tadiri, Scott, & Fussmann, 2016), gobies (Van Damme & Ollevier, 1994) and coral reef fish (Sikkel, Fuller, & Hunte, 2000). One reason for the sex difference in this system could be that males and females in the field are differently exposed to the parasite cercariae. Our sampling time in early summer coincided with the breeding season of gobies (Hesthagen, 1977; Lindström, 1998; Nyman, 1953) when males are defending nests and eggs, and remain mostly stationary. On the other hand, females are actively swimming around in search of spawning and feeding opportunities, which could increase their exposure to the parasite cercariae (see also Sikkel et al. (2000)). However, our exposure experiment showed that the sex difference in infections is not merely due to behavioral differences, but that females are also more susceptible to infection than males. This is interesting as it contradicts with the general trend of males being more susceptible to parasite infections because of sex hormones that can suppress immune function (Klein, 2004). The reason why such a sex-specific infection pattern was observed only in Vargskär is currently unclear, but may be related to differences between the habitat types. For example, the sampling sites Station and Långholmen are more sheltered compared to the more exposed Vargskär, although this would suggest lower rather than...
**TABLE 2**  Prevalence (% fish infected) and mean abundance (number of parasites per fish ± SE) of the parasite taxa detected in common gobies captured from three locations in the Northern Baltic Sea in June 2014–October 2015

| Parasite           | June 2014 | June 2015 | October 2015 |
|--------------------|-----------|-----------|--------------|
|                    | Station   | Långholmen| Vargskär     | Station | Långholmen| Vargskär     | Station | Långholmen| Vargskär     |
|                    | n = 30    | n = 24    | n = 0        | n = 30   | n = 30    | n = 0        | n = 0    | n = 24    | n = 30        |
| Trichodina sp.     | 100       | 95.8      | 96.7         | 100      | 15.3 ± 2.9| 0.0          | 60.0     | 0.9 ± 0.2  | 0.0           |
|                    | 18.0 ± 5.7| 32 ± 0.5  | 16.8 ± 4.2   | 0.0      | 0.0       | 0.0          | 0.0      | 0.0       | 0.0           |
| Gyrodactylus sp.   | 3.3       | 50.0      | 16.7         | 3.3      | 0.03 ± 0.03| 0.00 ± 0.03 | 0.0      | 0.07 ± 0.05| 0.07 ± 0.05   |
|                    | 0.03 ± 0.03| 0.7 ± 0.2 | 0.5 ± 0.3    | 0.00 ± 0.03| 0.03 ± 0.03| 0.00 ± 0.03 | 0.0      | 0.0       | 0.0           |
| Diplostomum spp.   | 0.0       | 0.0       | 3.3          | 0.03 ± 0.03| 0.3 ± 0.1 | 0.00 ± 0.03 | 83       | 13.3 ± 3.5 | 0.1 ± 0.1     |
|                    | 0.0       | 0.0       | 3.3          | 0.03 ± 0.03| 0.3 ± 0.1 | 0.00 ± 0.03 | 83       | 13.3 ± 3.5 | 0.1 ± 0.1     |
| Diplostomum baeri  | 0.0       | 0.0       | 0.0          | 0.0      | 0.0       | 0.0          | 0.0      | 0.0       | 0.0           |
|                    | 0.0       | 0.0       | 0.0          | 0.0      | 0.0       | 0.0          | 0.0      | 0.0       | 0.0           |
| Apatemon sp.       | 3.3       | 83        | 30           | 13.3     | 0.1 ± 0.1 | 0.1 ± 0.1    | 83       | 13.3 ± 3.5 | 0.1 ± 0.1     |
|                    | 0.07 ± 0.07| 0.1 ± 0.1 | 0.3 ± 0.1    | 0.1 ± 0.1| 0.1 ± 0.1 | 0.1 ± 0.1    | 83       | 13.3 ± 3.5 | 0.1 ± 0.1     |
| Cryptocotyle sp.   | 16.7      | 58.3      | 63.3         | 76.6     | 13.1 ± 3.5| 3.7 ± 1.3    | 45.8     | 73.3 ± 3.6 | 3.9 ± 0.8     |
|                    | 0.5 ± 0.4 | 2.0 ± 0.6 | 4.2 ± 1.4    | 13.1 ± 3.5| 3.7 ± 1.3 | 3.9 ± 0.8    | 45.8     | 73.3 ± 3.6 | 3.9 ± 0.8     |
| Sphaerostomum sp.  | 3.3       | 0.0       | 0.0          | 0.0      | 0.0       | 0.0          | 0.0      | 0.0       | 0.0           |
|                    | 0.03 ± 0.03| 0.0       | 0.0          | 0.0      | 0.0       | 0.0          | 0.0      | 0.0       | 0.0           |
| Proteocephalus sp. | 3.3       | 0.0       | 0.0          | 13.3     | 0.1 ± 0.1 | 0.03 ± 0.03  | 0.0      | 3.3 ± 0.03 | 0.03 ± 0.03   |
|                    | 0.03 ± 0.03| 0.0       | 0.0          | 13.3     | 0.1 ± 0.1 | 0.03 ± 0.03  | 0.0      | 3.3 ± 0.03 | 0.03 ± 0.03   |
| Camallanus lacustris| 6.7       | 0.0       | 0.0          | 0.0      | 0.0       | 0.0          | 0.0      | 0.0       | 0.0           |
|                    | 0.07 ± 0.05| 0.0       | 0.0          | 0.0      | 0.0       | 0.0          | 0.0      | 0.0       | 0.0           |
| Unidentified nematode| 6.7      | 4.2       | 0.0          | 3.3      | 0.03 ± 0.03| 0.04 ± 0.04  | 42       | 0.04 ± 0.04| 0.0           |
|                    | 0.07 ± 0.05| 0.04 ± 0.04| 0.03 ± 0.03| 0.04 ± 0.04| 0.04 ± 0.04| 0.04 ± 0.04 | 42       | 0.04 ± 0.04| 0.0           |
| Echinorhynchus gadii| 0.0       | 0.0       | 0.0          | 0.0      | 0.0       | 0.0          | 0.0      | 0.0       | 0.0           |
|                    | 0.0       | 0.0       | 0.0          | 0.0      | 0.0       | 0.0          | 0.0      | 0.0       | 0.0           |
| Neoechinorhynchus rutili| 0.0     | 4.2       | 0.0          | 0.0      | 0.0       | 0.0          | 0.0      | 0.0       | 0.0           |
|                    | 0.04 ± 0.04| 0.06 ± 0.04| 0.03 ± 0.03| 0.04 ± 0.04| 0.04 ± 0.04| 0.04 ± 0.04 | 0.0      | 0.0       | 0.0           |
higher infection in the latter. On the other hand, the population size of seagulls, terns, and mergansers, the definitive hosts for many trematodes including Diplostomum, may be larger at Vargskär than the other sites (Lindström & Ranta, 1992). This could enhance parasite life cycles locally (Marcogliese et al., 2001). However, details of the differences in infection processes need further work.

Our data also show that just one to two worms infecting an eye lens of small fish, such as gobies, can severely harm the lens (i.e., cataracts covering the majority of lens were formed at very low infection). This is because the size of the parasite metacercariae (and the damage they inflict per lens volume) likely remain relatively constant regardless of the lens size, but the size of the eye lens increases with fish size. Thus, in larger fish species, cataract coverage typically increases linearly with the parasite abundance and tens of parasites per lens may be required for high cataract intensities (Karvonen & Seppälä, 2008; Karvonen et al., 2004a), whereas even a low-level infection is likely to severely impair the vision of a small fish (see also Owen, Barber, and Hart (1993)). During mating, female sand gobies visit several males (Forsgren, 1997; Lindström & Lehtonen, 2013) and base their mate choice on a range of visual cues (Forsgren, 1992; Lindström et al., 2006). As a consequence, attractive males reach high mating success compared to less attractive ones (Lindström & Seppä, 1996) and this nonrandom distribution of mating success results in sexual selection (Andersson, 1994; Emlen & Oring, 1977). This process, however, can be potentially affected by impaired visual ability of females. For example, decreased water clarity interferes with visual abilities of females and this has been invoked as an explanation for weakened sexual selection in turbid waters (Järvenpää & Lindström, 2004; Seehausen et al., 1997). Similarly, if the visual ability of females was hampered by Diplostomum, this could affect the way females can judge males and express their mating preferences, as they would not be able to detect male mating signals. Consequently, mating systems in areas of high infection risk for females could become more random and result in weakened sexual selection. Such a process could easily create spatial variation in the intensity of sexual selection and may contribute to preserving genetic variation in male secondary sexual traits. However, female preferences are also based on the quality of male parental care (Lindström et al., 2006; Pampoule, Lindström, & St. Mary, 2004). Thus, it is possible that females in areas of high infection, on average, select males exhibiting lower quality care, which then results in lower offspring production. While our data are suggestive of the potential for such parasite-induced changes in mate choice, these questions need to be tackled experimentally.

To conclude, spatiotemporal variation in parasitism among populations, as well as that between sexes, can determine to what extent host populations suffer fitness consequences of infections in terms of decreased survival and reproductive success. Several earlier studies have suggested that sexual characteristics used in mate choice can advertise resistance of an individual toward parasite infections. However, the alternative that parasite infections could influence this process by impairing the visual ability of one sex to perceive such characteristics is virtually unexplored. Our data show the potential for such changes in mate choice in small-sized fish such as gobies, where visual abilities of females perceiving signals from males could deteriorate at very low infection intensities. However, we also suggest that such effects are likely to be different among host populations experiencing different levels of parasitism. Overall, relationships between spatiotemporal variation in infections, gender-biased parasitism and mate choice form an interesting field for further experimental research.

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CONFLICT OF INTEREST

The authors declare that they have no conflict of interest.

AUTHOR CONTRIBUTIONS

AK and KL conceived, designed, and performed the experiments. AK analyzed the data. AK and KL wrote the manuscript.

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