Epigenome-wide association of neonatal methylation and trimester-specific prenatal PM$_{2.5}$ exposure

Milan N. Parikh, Cole Brokamp, Erika Rasnick, Lili Ding, Tesfaye B. Mersha, Katherine Bowers, Alonzo T. Folger

**Background:** Exposure to particulate matter with an aerodynamic diameter smaller than 2.5 microns (PM$_{2.5}$) can affect birth outcomes through physiological pathways such as inflammation. One potential way PM$_{2.5}$ affects physiology could be through altering DNA methylation (DNAm). Considering that exposures during specific windows of gestation may have unique effects on DNAm, we hypothesized a timing-specific association between PM$_{2.5}$ exposure during pregnancy and DNAm in the neonatal epithelial-cell epigenome.

**Methods:** After collecting salivary samples from a cohort of 91 neonates, DNAm was assessed at over 850,000 cytosine-guanine dinucleotide (CpG) methylation sites on the epigenome using the MethylationEPIC array. Daily ambient PM$_{2.5}$ concentrations were estimated based on the mother’s address of primary residence during pregnancy. PM$_{2.5}$ was averaged over the first two trimesters, separately and combined, and tested for association with DNAm through an epigenome-wide association (EWA) analysis. For each EWA, false discovery rate (FDR)-corrected $P < 0.05$ constituted a significant finding and every CpG site with uncorrected $P < 0.0001$ was selected to undergo pathway and network analysis to identify molecular functions enriched by them.

**Results:** Our analysis showed that cg18705808 was associated with the combined average of PM$_{2.5}$. Pathway and network analysis revealed little similarity between the first two trimesters. Previous studies reported that TMEM184A, the gene regulated by cg18705808, has a putative role in inflammatory pathways.

**Conclusions:** The differences in pathway and network analyses could potentially indicate trimester-specific effects of PM$_{2.5}$ on DNAm. Further analysis with greater temporal resolution would be valuable to fully characterize the effect of PM$_{2.5}$ on DNAm and child development.

**Keywords:** DNA methylation; Fine particulate matter; Prenatal exposure

**Introduction**

A wide variety of environmental exposures can affect fetal development. One exposure of interest is ambient air pollution, specifically particulate matter with an aerodynamic diameter less than 2.5 microns (PM$_{2.5}$). Increased ambient PM$_{2.5}$ has adverse effects on human health, including mortality and morbidity related to respiratory, cardiovascular, and neurobehavioral health, through activation of inflammatory pathways. Neonatal outcomes are also negatively impacted by PM$_{2.5}$, with high exposures during fetal development having associations with low birth weights, abnormal lung development, and increased risk of preterm birth. Increased PM$_{2.5}$ exposures later in childhood have been associated with increased pediatric psychiatric emergency department utilization and increased risk for childhood cancer. While the association between PM$_{2.5}$ and these health outcomes is well established, the underlying mechanisms remain unclear. One potential pathway linking PM$_{2.5}$ exposure to long-term health outcomes is DNA methylation (DNAm), an epigenetic mechanism that is responsive to the environment and impacts health through changes in gene regulation.

DNA methylation is the process in which methyltransferases bind methyl groups to cytosine bases to form 5-methylcytosines, which can have a regulatory function by silencing gene expression. This study expands on previous research by presenting evidence for an effect of prenatal exposure to particulate matter with an aerodynamic diameter less than 2.5 microns (PM$_{2.5}$) on neonatal DNA methylation (DNAm) that varies based on the timing of the exposure. If these findings can be validated with larger sample sizes and higher temporal resolution, they may provide information on windows during which neonatal DNAm is sensitive to change by prenatal PM$_{2.5}$ exposure. We have also identified that PM$_{2.5}$ exposure is associated with DNAm at a locus within the body of gene TMEM184A, which has been implicated in inflammatory pathways.

What this study adds

This study expands on previous research by presenting evidence for an effect of prenatal exposure to particulate matter with an aerodynamic diameter less than 2.5 microns (PM$_{2.5}$) on neonatal DNA methylation (DNAm) that varies based on the timing of the exposure. If these findings can be validated with larger sample sizes and higher temporal resolution, they may provide information on windows during which neonatal DNAm is sensitive to change by prenatal PM$_{2.5}$ exposure. We have also identified that PM$_{2.5}$ exposure is associated with DNAm at a locus within the body of gene TMEM184A, which has been implicated in inflammatory pathways.
transcription.\textsuperscript{15} Cytosine-guanine dinucleotides (CpGs) are the most frequent sites of DNAm, and changes in DNAm at these sites can have functional consequences to biological processes and, therefore, health.\textsuperscript{16,17} DNAm, which is present in the genome from conception, is also mutable in the face of environmental factors, meaning that the transcription of the genome can change in response to exposures in the external and fetal environments.\textsuperscript{16,17} Therefore, identifying exposures that are associated with DNAm changes could suggest a mechanism for short- or long-term encoding of those exposures into the epigenome. In fact, altered DNAm associated with PM\textsubscript{2.5} levels has been identified in specific genes in several populations.\textsuperscript{24–26} As these methylation changes to the DNA could lead to adverse health effects, characterizing the relationship between exposures like fine particulate matter air pollution and DNAm can be a valuable step towards better understanding health and disease.

Exposure to exogenous environmental factors during pregnancy has been associated with DNAm changes at birth. For example, differing levels of socioeconomic status (SES) and community deprivation in the region of a child’s birth were associated with differential methylation of the infant’s epigenome at birth.\textsuperscript{27–29} In studies of air pollution, increased exposure to PM\textsubscript{2.5} during pregnancy was associated with a decrease in global DNAm as well as a decrease in methylation of specific genes in placental DNA at birth.\textsuperscript{29,30} Changes in DNAm associated with prenatal exposure to PM\textsubscript{2.5} can even be detected in salivary samples as late as at 15 years of age.\textsuperscript{31} Of particular note is a study utilizing a large multisite cohort that identified 14 CpG sites at which infant DNAm was associated with prenatal PM\textsubscript{2.5} exposure averaged across pregnancy.\textsuperscript{32} While averaging exposures across pregnancy does provide a robust and powerful marker of total PM\textsubscript{2.5} exposure during pregnancy, it is important to consider that the impact of an exposure on a specific health outcome may depend on the timing of that exposure.\textsuperscript{16,30} Given the sequential progression of anatomical and physiological structures during fetal development, exposures during specific windows of gestation may have unique effects on DNAm, depending on the phase of development occurring during that window.\textsuperscript{13,34} We hypothesized a timing-specific association between PM\textsubscript{2.5} exposure during pregnancy and differential methylation at specific CpG sites in the neonatal epigenome.

Our objectives were to (1) conduct an epigenome-wide association study (EWAS) and identify differentially methylated CpG sites in neonates that have DNAm values related to ambient PM\textsubscript{2.5} exposure levels during three different windows during pregnancy and (2) determine the biological networks and pathways enriched by the top CpG sites from the EWAS through pathway and network analysis to characterize potential downstream physiological effects of PM\textsubscript{2.5} exposure during the three windows of pregnancy.

**Methods**

**Participants**

This study combined participants from two related studies that enrolled from the same sampling frame, an ongoing home visiting program in Cincinnati, Ohio called Every Child Succeeds. Inclusion criteria for the program included that mothers were living in poverty, assessed through metrics such as income under 300% of the federal poverty level or receiving Medicaid.\textsuperscript{15} The Pregnancy and Infant Development (PRIDE) study was initiated as a pilot study in 2014, enrolling 53 mother-infant dyads, and was restarted in 2018 with a goal of enrolling 375 additional pairs.\textsuperscript{35,36} The PRIDE study enrolls pregnant women and follows the offspring postnatally at 1, 4, 12, and 18 months (follow-up in the pilot study was 1 month). Additional dyads were participants in a BioBank study (ECSBio) that collected specimens at 1 month and obtained similar exposure data. Pregnant women were enrolled between 2015 and 2019, and 93 have been followed through 1 month postnatal. The studies were approved by the Cincinnati Children’s Hospital Medical Center Institutional Review Board (IRB approval numbers: 2019-0808 and 2019-0588); all participants provided written informed consent.

**Measures**

Maternal demographics, including age and race, were collected at the first study visit (prenatal visit for PRIDE and 1-month postnatal for ECSBio). Additionally, the primary address of residence during pregnancy was collected as part of the study.

**Air pollution exposure assessment**

We geocoded each dyad’s primary residential address using a previously validated custom address range geocoder.\textsuperscript{37} Daily average PM\textsubscript{2.5} concentrations were estimated at a resolution of 0.75 sq km using a previously validated machine learning model that used data on regional PM\textsubscript{2.5} concentrations and emissions sources, planetary boundary layer height, wind speed, air temperature, and other spatiotemporal characteristics trained on PM\textsubscript{2.5} concentrations measured by the Environmental Protection Agency (EPA).\textsuperscript{38} The cross validated $R^2$ for this model in our study region was 0.97 for annual exposures and 0.96 for monthly exposures. Daily predictions were averaged for each dyad’s first trimester, second trimester, and over both of the first two trimesters of pregnancy through a simple arithmetic mean. The end of the second trimester was selected as a convenient endpoint due to the variability in the timing of birth in the third trimester, which leads to different lengths of pregnancy. In utero and ex utero, air pollution exposures are highly distinct, so we chose not to continue the investigation into the third trimester where birth could occur at different times. This two-trimester average represents the cumulative exposure over this period, while the individual trimester averages consider two major windows of exposure.

**DNA methylation**

DNA methylation (DNAm) was measured using salivary samples at a postnatal visit between 3 and 6 weeks of life. A total of 10 buccal swabs were collected from each neonate by swabbing with a sponge on the inner cheek until saturated and alternating between samples designated for DNA extraction (stored in lysate solution) and for cell spinning (stored in phosphate-buffered saline). The DNAgenotek OGR-250 infant saliva collection kits were used to collect cells for DNA extraction in the same manner as Folger et al.\textsuperscript{36} Genomic DNA went through sodium bisulfite (BS) conversion and was assayed using microarray technology in the Genomics, Epigenomics and Sequencing Core at the University of Cincinnati.\textsuperscript{22,39} BS-converted DNA was hybridized to the Illumina Infinium MethylationEPIC BeadChip, which assesses DNAm states at over 850,000 CpG sites in varied regions of the genome.

**Array processing**

Array processing was performed in R (R Core Team; Vienna, Austria), specifically using the minfi package (version 1.36.0) to import methylation data and estimate DNAm intensity.\textsuperscript{40} First, raw signal intensities were read into R, then quality control was performed at the sample level using kernel density plots of beta values and Illumina controls. Prediction of sex from methylation patterns was used to identify discrepancies between predicted and reported sex indicative of poor sample quality. The quality of the methylation arrays were evaluated in R (R Core Team; Vienna, Austria) using the ewastools package (version 1.7) to assess illumina’s 17 control metrics.\textsuperscript{41} All samples were within the limits for the Illumina controls.

CpG probes and sites also went through quality control. Probes/ sites were excluded from analysis if they (1) had a detection $P >$
the genes were included in the network not by chance was used to generate networks. The probability that Input CpG sites were mapped to genes with both indirect and network interactions to generate putative pathways and networks.

Netherlands) that uses an expert level curated database of pathways and networks implicated by the genes in which those sites resided.

Results

Exclusions and sample characteristics

For this analysis, data were collected from 56 subjects in PRIDE and from 37 in ECSBio. Of these original 93 subjects, 91 were used in the final analysis. One subject was excluded due to an address that could not be geocoded precisely enough to assess air pollution. A second subject was excluded due to a mismatch between true sex and sex predicted from methylation patterns indicating unreliable methylation data. Table 1 details summary statistics for our sample of 91 mother-infant dyads. PM2.5 exposure, measured in μg/m³, averaged during trimester 1 had median 9.06, interquartile range (IQR) 8.47–10.36, and range 7.58–11.50. Exposure averaged during trimester 2 had median 8.15, IQR 7.89–8.75, and range 7.27–11.36, and exposure averaged over both trimesters 1 and 2 had median 8.71, IQR 8.45–9.30, and range 7.98–10.15.

Quality control

After probe-level and site-level quality control, 3,837 CpG sites were excluded for low bead counts, 221,486 were excluded with detection \( P > 0.01 \), and 55,081 were excluded due to locations on sex chromosomes, locations at or near SNP loci, or potential cross-hybridization. Following exclusions, a total of 585,604 CpG sites were available for epigenome-wide associations. The QQ plots from the EWA with trimester 1 (eFigure 1; http://links.lww.com/EE/A203), trimester 2 (see eFigure 3; http://links.lww.com/EE/A203), and the combined period (eFigure 5; http://links.lww.com/EE/A203) did not show large deviations between expected and observed \( P \)-values.

Average prenatal \( PM_{2.5} \) exposure

None of the CpG sites included in the analysis of \( PM_{2.5} \) averaged over trimester 1 were found to have DNAm significantly associated with average \( PM_{2.5} \) exposure in the first trimester after FDR correction (eFigure 2; http://links.lww.com/EE/A203).

Characteristics Included subjects (n = 91)

| Characteristics | Included subjects (n = 91) |
|-----------------|---------------------------|
| Maternal factors |                           |
| Enrollment age (years); mean (SD) | 22.6 (4.1) |
| Race, n (%) | 50 (54.9) |
| White | 33 (36.3) |
| Other | 8 (8.8) |
| Child factors |                           |
| Sex, n (%) | 42 (46.2) |
| Female | 49 (53.8) |
| Gestational age at birth (weeks); mean (SD) | 38.9 (1.6) |
PPP1CB, TMEM184A, GEMIN5, MBD3, CTDPI, KCNQ4, PLIN4, MBP, and KCNQ1. No genes from other EWAs overlapped with genes housing sites with $P < 0.0001$ in the trimester 2 EWA. Summaries of major pathways and networks identified by IPA pathway and network analysis are detailed in Table 3.

**Discussion**

In this study, DNAm at one CpG site in the body of the TMEM184A gene was significantly associated with PM$_{2.5}$ averaged over the first two trimesters. No significant associations were observed between exposure averaged over trimesters 1 or 2 and DNAm. There was no overlap in the genes housing the 10 CpG sites with the lowest $P$ values in trimester 1 and trimester 2. Additionally, IPA revealed very different pathways and networks implicated by the top CpG sites in each trimester. The TMEM184A gene encodes for the transmembrane protein 184A (TMEM184A). This transmembrane protein is found in vascular smooth muscle cells and endothelial cells and is a key receptor for heparin in vascular cells. TMEM184A plays an essential role in heparin-induced anti-inflammatory activity in vascular endothelial cells. Although it is unclear whether this hypermethylation is mechanistically responsible for PM$_{2.5}$-induced inflammation, the observed association between increased PM$_{2.5}$ exposure and cg18705808 DNAm is consistent with previously established links between PM$_{2.5}$ and inflammation.

The observation that IPA produced different results when given top CpG sites from different periods could suggest that PM$_{2.5}$ exposure during pregnancy has unique downstream physiological effects depending on the timing of exposure. A different set of pathways and networks were associated with trimester 1 PM$_{2.5}$ exposure compared with trimester 2. This could suggest a timing-specific association between PM$_{2.5}$ exposure during pregnancy and DNAm in offspring buccal epithelial cells. However, using IPA with CpG sites that do not pass the significance threshold is an imperfect investigation. There is a chance that the differences seen are simply a result of chance. An analysis of specific CpG sites of interest with better temporal resolution for PM$_{2.5}$ exposure would help to elucidate this relationship.

In a similar analysis of 1,551 participants, Gruzieva et al$^{32}$ identified 14 CpG sites associated with average prenatal PM$_{2.5}$ exposure across pregnancy, a corollary to cumulative exposure similar to our two trimester average. The CpG sites found significant by their analysis did not overlap with any of the top sites from our study, in part due to their measurement of cord blood DNAm instead of buccal epithelial cell DNAm. Their notably larger sample size lends them greater power, explaining their ability to identify so many significant associations. However, our study also included trimester-specific windows of exposure while Gruzieva et al$^{32}$ examined only a cumulative exposure model. Considering the association between prenatal PM$_{2.5}$ and infant DNAm at specific CpG sites with exposure occurring during specific trimesters of pregnancy affords us a

---

**Figure 1.** Manhattan plots for the association between DNA methylation at each studied CpG site and average PM$_{2.5}$ over trimesters 1 and 2 of pregnancy adjusted for specified covariates. The vertical axis plots the negative log$_{10}$ $P$ value for each association. The Bonferroni cutoff for significance is denoted by the horizontal red line.
pollution concentrations will allow us to make inferences on potential primary prevention strategies and allowed us to study the impact of air pollution on the epigenome without conducting expensive and timely personal sampling. Lastly, averaging PM$_{2.5}$ across a trimester may not capture fluctuations of exposure over time, which may have biologic relevance. It is possible that exposure to PM$_{2.5}$ is only associated with DNAm changes during short windows of time or that exposure closer to birth during the third trimester has a greater effect on DNAm. Those effects may be missed when averaging the exposure over the trimester. Additionally, this method is not equipped to consider that PM$_{2.5}$ concentrations during the selected window may be associated with concentrations outside the window and that the effect seen is truly caused by those concentrations that are not included in the average. Future analyses should use statistical methods that can utilize daily PM$_{2.5}$ measurements to provide greater temporal resolution.

In summary, our EWA identified an association between PM$_{2.5}$ exposure averaged over the first two trimesters of pregnancy with the hypermethylation of cg18705808, a CpG site

| Table 2. Ten CpG sites with the lowest $P$ values from each EWA |
|---------------------------------------------------------------|
| **Trimester 1**                                               |
| **CpG site** | Chr | Gene | Location | M coefficient | $P$ | $\beta$ coefficient (95% CI) |
|---------------|-----|------|----------|----------------|-----|----------------------------|
| cg14305641$^a$ | 19  | MBP3 | 3' UTR   | 0.792          | 4.62 × 10$^{-7}$ | 0.015 (0.008, 0.022) |
| cg16668903    | 5   | SNX24| Body     | 0.714          | 6.83 × 10$^{-7}$ | 0.013 (0.007, 0.018) |
| cg10940724    | 9   | DPP7 | TSS1500  | 1.615          | 1.82 × 10$^{-5}$ | 0.142 (0.086, 0.198) |
| cg10419550    | 8   | Intergenic | Intergenic | 1.094          | 3.17 × 10$^{-5}$ | 0.175 (0.109, 0.240) |
| cg13894539$^b$ | 19  | PLIN4| TSS1500  | 0.949          | 3.92 × 10$^{-5}$ | 0.211 (0.134, 0.289) |
| cg16180217    | 3   | DDBL2| TSS200   | 0.598          | 4.56 × 10$^{-5}$ | 0.132 (0.076, 0.186) |
| cg08672630    | 6   | Intergenic | Intergenic | 1.758          | 4.72 × 10$^{-5}$ | 0.404 (0.241, 0.568) |
| cg25799969$^b$ | 2   | PPP1CB| 5' UTR, 1st Exon | 1.758       | 6.38 × 10$^{-6}$ | 0.167 (0.088, 0.246) |
| cg07309114    | 17  | NPEPPS| Body     | 1.929          | 8.05 × 10$^{-6}$ | 0.198 (0.111, 0.286) |
| cg02330310    | 1   | YTHDF2| Body     | 1.414          | 8.59 × 10$^{-6}$ | 0.343 (0.199, 0.486) |
| **Trimester 2**                                               |
| **CpG site** | Chr | Gene | Location | M coefficient | $P$ | $\beta$ coefficient (95% CI) |
|---------------|-----|------|----------|----------------|-----|----------------------------|
| cg11845050    | 22  | PHKAP2| TSS200   | 0.766          | 1.30 × 10$^{-7}$ | 0.033 (0.022, 0.044) |
| cg13804427    | 10  | PROSER2| 5' UTR   | 0.931          | 6.73 × 10$^{-5}$ | 0.056 (0.037, 0.075) |
| cg02397114    | 17  | TMEM102| TSS200  | 0.673          | 7.72 × 10$^{-6}$ | 0.023 (0.013, 0.032) |
| cg10180496    | 4   | UBE203| TSS1500, 5' UTR, TSS200, Body | 1.450       | 8.83 × 10$^{-5}$ | 0.101 (0.061, 0.142) |
| cg10156714    | 20  | LOC103377348| TSS1500 | 0.635          | 8.99 × 10$^{-5}$ | 0.024 (0.016, 0.032) |
| cg25967419    | 6   | KLC4  | TSS1500, TSS200 | 1.872       | 1.09 × 10$^{-5}$ | 0.045 (0.029, 0.061) |
| cg10149123    | 16  | C16or91| Body    | -2.231         | 1.28 × 10$^{-5}$ | -0.277 (-0.377, -0.177) |
| cg16493531    | 6   | FLOT1 | TSS1500  | -1.979         | 1.45 × 10$^{-5}$ | -0.424 (-0.580, -0.267) |
| cg23528705    | 7   | ERP3  | Body     | -1.980         | 1.45 × 10$^{-5}$ | -0.425 (-0.582, -0.270) |
| cg22979231    | 4   | UNEX  | Body     | 1.600          | 1.50 × 10$^{-5}$ | 0.075 (0.059, 0.098) |
| **Trimesters 1 and 2 combined**                              |
| **CpG site** | Chr | Gene | Location | M coefficient | $P$ | $\beta$ coefficient (95% CI) |
|---------------|-----|------|----------|----------------|-----|----------------------------|
| cg18705808$^a$ | 7   | TMEM184A| Body    | 1.375          | 1.93 × 10$^{-6}$ | 0.025 (0.015, 0.036) |
| cg13894539$^b$ | 19  | PLIN4 | TSS1500  | 11.017         | 5.00 × 10$^{-7}$ | 1.328 (0.762, 1.933) |
| cg13852093    | 18  | MBP   | Body     | 16.052         | 5.20 × 10$^{-3}$ | 2.189 (1.491, 2.877) |
| cg17706097    | 10  | Intergenic | Intergenic | 1.949          | 1.97 × 10$^{-5}$ | 0.312 (0.182, 0.443) |
| cg11931463    | 3   | Intergenic | Intergenic | 1.981          | 4.02 × 10$^{-5}$ | 0.411 (0.239, 0.584) |
| cg09308091    | 10  | ZMIZ1 | Body     | 1.572          | 4.62 × 10$^{-5}$ | 0.301 (0.178, 0.424) |
| cg14305641$^a$ | 19  | MBP3  | 3' UTR   | 3.704          | 4.77 × 10$^{-5}$ | 0.192 (0.096, 0.288) |
| cg25799969$^b$ | 2   | PPP1CB| 5' UTR, 1st Exon | 3.657       | 7.03 × 10$^{-6}$ | 0.847 (0.540, 1.155) |
| cg2711750S    | 2   | OSP1  | 3' UTR   | 3.479          | 7.55 × 10$^{-6}$ | 0.181 (0.098, 0.265) |
| cg26020805    | 18  | CDTP1 | Body     | 2.950          | 7.94 × 10$^{-6}$ | 0.240 (0.147, 0.351) |

$^a$Sites in both top 10 of trimester 1 and combined EWAs.
$^b$Site found to be significant after FDR correction.
$^c$Regression parameter from EWAs on DNAm $\beta$-value corresponding an increase in 10 $\mu$g/m$^3$ of PM$_{2.5}$
$^d$Pathway and Network Analysis.

- **CI indicates confidence interval.**

more detailed perspective on the effect of PM$_{2.5}$ exposure on development. It is true that we did not find results in the individual trimester analyses, which passed the FDR threshold for significance, but a lack of significant findings does not strictly mean that there are no differences between the two analyses. The fact that the top sites are vastly different between the first and second trimester EWAs suggests to us that differences in the effect of PM$_{2.5}$ exposure between trimesters may exist, and this possibility deserves further exploration.

Our study had some limitations. First, the small sample size limited our power and prevented more complex analyses. It also limits the external generalizability of our findings when considering populations with different PM$_{2.5}$ concentrations and composition. Although ambient air pollution exposures used here do not represent personal exposures, it was the most appropriate measure that could be used to answer our research question without introducing confounding by personal activities, behaviors, and characteristics. Studying ambient versus personal air pollution concentrations will allow us to make inferences on...
located within the gene TMEM184A. The associated protein TMEM184A is a vital receptor in a heparin-induced anti-inflammatory pathway, a finding consistent with the links between PM$_{2.5}$ and inflammation. Pathway and network analysis revealed largely different biological processes implicated by PM$_{2.5}$ exposure in trimester 1 versus in trimester 2, which could suggest a timing-specific effect of prenatal PM$_{2.5}$ exposure on infant DNA methylation. More research needs to be done with a larger sample size and utilizing an approach with greater temporal resolution. To explore this proposed timing-specific effect in combination with established effects of long-term cumulative exposures.

Conflicts of interest statement

The authors declare that they have no conflicts of interest with regard to the content of this report. The results reported herein correspond to specific aims of grant R56MD013006 to investigator A.T.F. and K.B. from the National Institute of Minority Health and Health Disparities. T.B.M. was supported by the National Institutes of Health (NIH) R01 HL132344 and R01 HG011441 grants. This work was also supported by grants from the Cincinnati Children’s Hospital Medical Center including Academic Research Committee and Trustee awards to K.B. and A.T.F.

Acknowledgments

We would like to thank the many families and facilitators involved in the PRIDE and ECS BioBank studies for their contributions to the data used in this study.

References

1. Schwartz J, Dockery DW, Neas LM. Is daily mortality associated specifically with fine particles? J Air Waste Manag Assoc. 1996;46:927–939.
2. Fu H, Liu X, Li W, et al. PM2.5 exposure induces inflammatory response in macrophages via the TLR4/COX-2/NF-κB pathway. Inflammation. 2020;43:1948–1958.
3. Block ML, Calderón-Garcidueñas L. Air pollution: mechanisms of neurodevelopmental and CNS disease. Trends Neurosci. 2009;32:506–516.
4. Costa LG, Cole TB, Coburn J, et al. Neurotoxins are in the air: convergence of human, animal, and in vitro studies on the effects of air pollution on the brain. Bale A, ed. Biomed Res Int. 2014;2014:736385.
5. Allen JL, Klocke C, Morris-Schaffer K, et al. Cognitive effects of air pollution exposures and potential mechanistic underpinnings. Curr Environ Heal Reports. 2017;4:180–191.
6. Prunicki M, Stell L, Dinakarpandian D, et al. Exposure to NO$_x$, CO, and PM$_{2.5}$ is linked to regional DNA methylation differences in asthma. Clin Epigenetics. 2018;10:2.
7. Bell ML, Belanger K, Ebius K, et al. Prenatal exposure to fine particulate matter and birth weight: variations by particulate constituents and sources. Epidemiology. 2010;21:884–891.
8. Yue H, Ji X, Li G, et al. Maternal exposure to PM$_{2.5}$ affects fetal lung development at sensitive windows. Environ Sci Technol. 2020;54:316–324.
9. Liu S, Krewski D, Shi Y, et al. Association between maternal exposure to ambient air pollutants during pregnancy and fetal growth restriction. J Expo Sci Environ Epidemiol. 2007;17:426–432.
10. Shah PS, Balkhair T. Air pollution and birth outcomes: a systematic review. Environ Int. 2011;37:498–516.
11. Dadvand P, Parker J, Bell ML, et al. Maternal exposure to particulate air pollution and term birth weight: a multi-country evaluation of effect and heterogeneity. Environ Health Perspect. 2013;121:267–273.
12. Brokamp C, Stawnn JR, Beck AF, et al. Pediatric psychiatric emergency department utilization and fine particulate matter: a case–crossover study. Environ Health Perspect. 2019;127:97006.
13. Avisine E, Belair M-A, Do MT, et al. Maternal exposure to ambient air pollution and risk of early childhood cancers: a population-based study in Ontario, Canada. Environ Int. 2017;100:139–147.
14. Cao-Lei L, de Rooij SR, King S, et al. Prenatal stress and epigenetics. Neurosci Biobehav Rev. 2020;117:198–210.
15. Michels KB. Epigenetic Epidemiology. 1st ed. Springer; 2012.
16. Kundakovic M, Jaric I. The epigenetic link between prenatal adverse environments and neurodevelopmental disorders. Genes (Basel). 2017;8:104.
17. Jones DE, Park JS, Gamby K, et al. Mental health epigenetics: a primer with implications for counselors. Prof Couns. 2021;11:102–121.
18. Keverne EB, Curley JP. Epigenetics, brain evolution and behaviour. Front Neuroendocrinol. 2008;29:398–412.
19. Baccarelli A, Wright RO, Bollati V, et al. Rapid DNA methylation changes after exposure to traffic particles. Am J Respir Crit Care Med. 2012;179:572–578.
20. Conradt E, Hawes K, Guerin D, et al. TET1 methylation is associated with childhood asthma and DNA methylome variation in childhood. Dev Psychopathol. 2018;30:891–903.
21. Sommenni HK, Zhang X, Bagnini Myers JM, et al. Ten-eleven translocation 1 (TET1) methylation is associated with childhood asthma and traffic-related air pollution. J Allergy Clin Immunol. 2016;137:797–805.e5.
22. Szyf M, Tang Y-Y, Hill KG, et al. The dynamic epigenome and its implications for behavioral interventions: a role for epigenetics to inform disorder prevention and health promotion. Transl Behav Med. 2016;6:55–62.
23. Wu Y, Qie R, Cheng M, et al. Air pollution and DNA methylation in adults: a systematic review and meta-analysis of observational studies. Environ Pollut. 2021;284:117152.
24. Wang C, Chen R, Cai J, et al. Personal exposure to fine particulate matter and blood pressure: a role of angiotensin converting enzyme and its DNA methylation. Environ Int. 2016;94:661–666.

Table 3.

Top pathways and networks enriched by genes regulated by top CpG sites using IPA

| Trimester 1 | -log10(P) | Top canonical pathways |
|-------------|----------|------------------------|
| 5-aminoimidazole ribonucleotide biosynthesis I | 2.62 | Human embryonic stem cell pluripotency |
| Purine nucleotides de novo biosynthesis II | 2.06 | Circadian rhythm signaling |
| DNA methylation and transcriptional repression signaling | 1.56 | Complement system |
| Dermatan sulfate biosynthesis (late stages) | 1.43 | G protein signaling mediated by tubby |
| Chondroitin sulfate biosynthesis (late stages) | 1.41 | Top networks |
| Cancer, cardiac necrosis/cell death, cell death and survival | 23 | Cellular response to therapeutics, cell-to-cell signaling and interaction, nervous system development and function |
| Cancer, cellular compromise,organismal injury and abnormalities | 23 | Embryonic development, organismal development, gene expression |

| Trimester 2 | -log10(P) | Top canonical pathways |
|-------------|----------|------------------------|
| Iron homeostasis signaling pathway | 1.94 | Infectious diseases, cell-to-cell signaling and interaction, renal and urological system development and function |
| Human embryonic stem cell pluripotency | 1.79 | Cellular movement |
| Circadian rhythm signaling | 1.4 | Top networks |
| Complement system | 1.37 | Score |
| G protein signaling mediated by tubby | 1.3 | Top networks |
| Cell death and survival, connective tissue development and function, cellular movement | 26 | DNA damage and repair, protein synthesis, infectious diseases |
| Developmental disorder, hereditary disorder, ophthalmic disease | 2 |

Scores for top networks represent a ranking of the networks least likely to be generated by chance with scores over 2 denoting a 99% likelihood of not being generated by chance.

| Trimesters 1 and 2 | -log10(P) | Top canonical pathways |
|-------------------|----------|------------------------|
| Wnt/Ca$^{2+}$ pathway | 2.83 | Wnt signaling pathway |
| Wound healing signaling pathway | 2.59 | Insulin secretion signaling pathway |
| Pulmonary fibrosis idiopathic signaling pathway | 2.28 | Neurogenic pain signaling in dorsal horn neurons |
| Neurogenic pain signaling in dorsal horn neurons | 2.27 | Top networks |
| Lipid metabolism, small molecule biochemistry, cell cycle | 2.7 | Score |
| Cell death and survival, connective tissue development and function, cellular movement | 24 | RNA damage and repair, protein synthesis, infectious diseases |
| Developmental disorder, hereditary disorder, ophthalmic disease | 2 |

Parikh et al. • Environmental Epidemiology (2022) 6:e227
26. Salam MT, Byun HM, Lurmann F, et al. Genetic and epigenetic variations in inducible nitric oxide synthase promoter, particulate pollution, and exhaled nitric oxide levels in children. J Allergy Clin Immunol. 2012;129:232–239.e1–e7.
27. Laubach ZM, Perng W, Cardenas A, et al. Socioeconomic status and DNA methylation from birth through mid-childhood: a prospective study in Project Viva. Epigenomics. 2019;11:1413–1427.
28. DeLano K, Folger AT, Ding L, et al. Associations between maternal community deprivation and infant DNA methylation of the SLC6A4 gene. Front Public Heal. 2020;8:537793.
29. Janssen BG, Godderis L, Pieters N, et al. Placental DNA hypomethylation in association with particulate air pollution in early life. Part Fibre Toxicol. 2013;10:1–11.
30. Saenen ND, Vrijens K, Janssen BG, et al. Lower placental leptin promoter methylation in association with fine particulate matter air pollution during pregnancy and placental nitrosative stress at birth in the ENVIRONAGE cohort. Environ Health Perspect. 2017;125:262–268.
31. Bakulski KM, Fisher JD, Dou JF, et al. Prenatal particulate matter exposure associated with saliva DNA methylation at age 15: applying cumulative DNA methylation scores as an exposure biomarker. Toxics. 2021;9:262.
32. Gruzieva O, Xu CJ, Yousefi P, et al. Prenatal particulate air pollution and DNA methylation in newborns: an epigenome-wide meta-analysis. Environ Health Perspect. 2019;127:57012.
33. Tobi EW, Lumey LH, Talens RP, et al. DNA methylation differences after exposure to prenatal famine are common and timing- and sex-specific. Hum Mol Genet. 2009;18:4046–4053.
34. Abraham E, Rousseaux S, Agier L, et al. Pregnancy exposure to atmospheric pollution and meteorological conditions and placental DNA methylation. Environ Int. 2018;118:334–347.
35. Bowers K, Ding L, Yolton K, et al. Pregnancy and Infant Development (PRIDE)—a preliminary observational study of maternal adversity and infant development. BMC Pediatr. 2021;21:452.
36. Folger AT, Ding L, Ji H, et al. Neonatal NR3C1 methylation and DNA methylation scores as an exposure biomarker. Toxics. 2021;9:262–268.
37. Brokamp CA. High resolution spatiotemporal fine particulate matter exposure assessment model for the contiguous United States. Environ Adv. 2022;7:100155.
38. Ji H, Zhang X, Oh S, et al. Dynamic transcriptional and epigenomic reprogramming from pediatric nasal epithelial cells to induced pluripotent stem cells. J Allergy Clin Immunol. 2015;135:236–244.
39. Ayre MJ, Jaffe AE, Corrada-Bravo H, et al. Minfi: a flexible and comprehensive Bioconductor package for the analysis of Illumina DNA methylation microarrays. Bioinformatics. 2014;30:1363–1369.
40. Heiss JA, Just AC. Identifying mislabeled and contaminated DNA methylation microarray data: an extended quality control toolset with examples from GEO. Clin Epigenetics. 2018;10:73.
41. Heiss JA, Just AC. Improved filtering of DNA methylation microarray data by detection p values and its impact on downstream analyses. Clin Epigenetics. 2019;11:15.
42. Prunicki M, Cauwenberghs N, Lee J, et al. Air pollution exposure is linked with methylation of immunoregulatory genes, altered immune cell subsets, and IFNγ. Clin Epigenetics. 2022;14:40.
43. Foussoux N, rest of the text continues...