Sources of error in open-path FTIR measurements of N$_2$O and CO$_2$ emitted from agricultural fields

Cheng-Hsien Lin$^1$, Richard H. Grant$^1$, Albert J. Heber$^2$, and Cliff T. Johnston$^{1,3}$

$^1$Department of Agronomy, Purdue University, West Lafayette, IN 47907, United States
$^2$Department of Agricultural and Biological Engineering, Purdue University, West Lafayette, IN 47907, United States
$^3$Department of Earth, Atmospheric and Planetary Sciences, Purdue University, West Lafayette, IN 47907

Correspondence to: Cheng-Hsien Lin (lin471@purdue.edu)

Abstract. Open-path Fourier transform infrared spectroscopy (OP-FTIR) is susceptible to environmental variables which can become sources of errors for gas quantification. In this study, we assessed the effects of water vapour, temperature, path length, and wind speed on the uncertainty of nitrous oxide (N$_2$O) and carbon dioxide (CO$_2$) concentrations derived from OP-FTIR spectra. The presence of water vapour resulted in underestimating N$_2$O in both lab (−3 %) and field (−12 %) experiments using a classical least squares (CLS) model when the reference and sample spectra were collected at the same temperature (i.e. 30 °C). Differences in temperature between the sample and reference spectra also underestimated N$_2$O concentrations due to temperature broadening and the increased interferences of water vapour in spectra of wet samples. Changes in path length resulted in a non-linear response of spectra and bias (e.g. N$_2$O and CO$_2$ concentrations were underestimated by 30 % and 7.5 %, respectively, at the optical path of 100 m using CLS models). For N$_2$O quantification, partial least squares (PLS) models were less sensitive than CLS to the influence of water vapour, temperature, and path length, and provided more accurate estimations. Uncertainties in the path-averaged concentrations increased in low wind conditions (< 2 m s$^{-1}$). This study identified the most common interferences that affect OP-FTIR measurements of N$_2$O and CO$_2$, which can serve as a quality assurance/control guide for current or future OP-FTIR users.

1 Introduction

Agriculture substantially contributes greenhouse gases (GHGs), mostly N$_2$O and CH$_4$, to the atmosphere (IPCC, 2007). In 2010, emissions led by agricultural activities (e.g. crop production and livestock management) were estimated between 5.2 and 5.8 Gt of CO$_2$ equivalent per year, accounting for 10-12 % global anthropogenic emissions (IPCC, 2014). Estimations of gas fluxes over an extended period (e.g. growing seasons) is complicated due to the dynamic and episodic nature of gas emissions and measurement complexities. The integrated uses of fast-response gas concentration sensors and micrometeorological techniques were developed to measure long-term gas fluxes continuously (Baldocchi, 2003; Denmead, 2008; Flesch et al., 2016). Open-path Fourier transform spectroscopy (OP-FTIR) is capable of measuring concentrations of multiple gases simultaneously with high temporal and spatial resolution through real-time measurements and path-averaged concentrations (Russwurm and Childers, 2002). OP-FTIR has been applied to measure GHGs, and other trace gases (e.g. NH$_3$)
emitted from agricultural fields (Childers et al., 2001a and 2001b; Bjorneberg et al., 2009; Flesch et al., 2016; Lam et al., 2017). Using OP-FTIR to quantify gas concentrations, however, is a complicated process. Each step in data collection and spectral analyses influences the accuracy and precision of gas quantifications, including spectral resolution, apodization function, choice of background (e.g. zero-path or synthetic backgrounds), and chemometric models (Russwurm and Childers, 1999; Griffiths and de Haseth, 2007; Hart and Griffiths, 2000; Hart et al., 2000). Also, OP-FTIR spectra are sensitive to ambient environmental conditions (e.g. humidity, air temperature, optical distance, and wind speed), which interferes with spectral analyses and quantification of gas concentrations (Griffiths and de Haseth, 2007; Muller et al., 1999; Shao et al., 2007 and 2010).

Water vapour is a major interference in FTIR-derived trace gas quantification due to its strong absorption features within the mid-infrared region (400-4000 cm⁻¹) (Russwurm and Childers, 1999; ASTM, 2013). For the interference-free spectra, a single absorption line can be easily isolated to calculate gas concentrations (i.e. univariate methods). Using this method, however, is challenging to adequately isolate the absorption feature of the gas of interest from water vapour (Muller et al., 1999; Briz et al., 2007). Multivariate methods have been proposed to improve gas quantification from spectral interferences including water vapour by selecting broad spectral windows of gases of interest (Haaland and Easterling, 1980; Haaland and Thomas, 1988; Hart and Griffiths, 1998 and 2000; Hart et al., 1999 and 2000; Muller et al., 1999). The most common method in commercial quantification packages is the classical least squares (CLS) quantitative model (Griffiths and de Haseth, 2007). Studies showed that the interference of water vapour was mitigated by either considering reference spectra of water vapour or through selection of appropriate spectral windows in CLS models (ASTM, 2013; Haaland and Easterling, 1980; Hart and Griffiths, 2000; Horrocks et al., 2001; Jiang et al., 2002; Du et al., 2004; Briz et al., 2007; Lin et al., 2019). Other studies, however, found that CLS models resulted in substantial quantification errors of the targeted gas under the interference of the non-targeted gases (mostly water vapour) even if the reference spectra of all gas species and the optimal spectral window were considered (Hart et al., 1999; Briz et al., 2007; Shao et al., 2010; Lin et al., 2019). As a result, the partial least squares (PLS) algorithm was proposed to improve the accuracy of gas quantification (Haaland and Thomas, 1988; Hart et al., 2000; Shao et al., 2010; Lin et al., 2019).

The features of gas rotation-vibrational absorption bands are temperature dependent (Lacome et al., 1984; Rothman et al., 2005). Ideally, sample and reference spectra should be collected at the same temperature to avoid temperature-related bias (Russwurm and Childers, 1999; ASTM, 2013). Training spectra for building quantitative models, such as CLS, were generally collected at the same temperature. The non-linear responses of spectral absorbance to significant diurnal variations in temperature will lead to errors in gas quantification (Russwurm and Phillips, 1999; de Castro et al., 2007; Smith et al., 2011). For continuous field measurements, it is time-consuming to create piecewise models to cover the entire ranges of 1) the path-integrated concentrations of gases of interest and interferences, and 2) temperature. A PLS model was developed to cover the
wide ranges of environmental variables, including concentrations, path lengths, humidity, and temperature (Bjorneberg et al., 2009; Griffiths et al., 2009; Shao et al., 2010 and 2011).

Besides changes in water vapour content and temperature, the experimental configuration and optical parameters also influence OP-FTIR spectra. Compared with extractive methods, one of the advantages of OP-FTIR is the ability to use different path lengths to measure gases from multiple sources (Russwurm and Childers, 1999; Bacsik et al., 2006). OP-FTIR measurements require electronic gain to fill the analogue-to-digital converter (A/DC) of the instrument. Consequently, short optical path lengths can over saturate the detector that introduces error in gas quantification (Bartoli et al., 1974; Chase, 1984; Griffiths and de Haseth, 2007). A long path length increases the sensitivity for gas quantification, but the increased length reduces the intensity of the incident signal and decreases the signal-to-noise (S/N) ratio (Griffith and Jamie, 2006; Nelson et al., 1996; Kosterev et al., 2008). Spectral resolution and apodization also affect the spectral linearity (Griffiths, 1994; Zhu and Griffiths, 1998; Russwurm and Phillips, 1999; Childers et al., 2002). Lower resolution spectra are incapable of resolving absorption features. Even though the apodized interferogram (IFG) can reduce noises (or spurious oscillations) of a single-beam (SB) spectrum converted by Fourier transformation, different apodization functions affect spectrum linearity differently. The non-linear relationship between absorbance and concentrations disobeys Beer-Lambert Law and leads to biases in gas quantification (Haaland, 1987; Russwurm and Phillips, 1999; Childers et al., 2001a).

The CLS algorithm is subject to the nonlinearity induced by ambient interferences (e.g. multiple overlapping components) or instrumentation setup (e.g. low spectral resolution) which can be overcame by the PLS-based quantitative models (Hart et al., 2000). Alternatively, the non-linear least squares (NLLS) regression function was also proposed to correct non-linear behaviour of the spectrum to concentrations and improve the quantitative accuracy (Smith et al., 2011; Griffith et al., 2012; Paton-Walsh et al., 2014; Phillips et al., 2019). The NLLS coupled with a forward modelling approach named the Multi-Atmospheric Layer Transmission (MALT) method can model a transmission spectrum from the HITRAN database by selecting specific parameters (including temperature, pressure, path length, resolution, and apodization function) to fit the sampled SB spectrum for gas quantification (Griffith, 1996). One of the benefits of the MALT is that a ‘background’ SB spectrum (the spectra without gases of interest) is not required in the procedure for quantification, which can avoid the background-induced errors (e.g. zero-path background or the zapped background shown in Lin et al, 2019). Nevertheless, it is challenging to consider other sources of variables in the field (e.g. severe spectral overlap or wind-induced variations) using the MALT which is built upon the HITRAN library. Instead, the PLS-based approach can be trained empirically and consider the local environmental variables in quantitative models for field gas measurements (ASTM, 2013; Griffith et al., 2009).

Many OP-FTIR field studies used either CLS or PLS as frameworks for gas measurements, and most of these studies minimized environmental interferences (e.g. water vapour or wind speed) by developing methods for spectral analyses and gas quantification (Hong and Cho, 2003; Hart et al., 1999 and 2000; Muller et al., 1999; Childers et al., 2002; Briz et al., 2007;
Shao et al., 2007; Griffiths et al., 2009; Shao et al., 2010; Lin et al., 2019). Briz et al. (2007) examined the difference in water vapour content (0.5 % vs. 2.5 %) on trace gas quantification but their study did not include N$_2$O. Shao et al. (2007) investigated the effect of wind on spectrometer vibrations and spectra qualities but did not address its influence on gas quantification. Temperature effect on spectral parameters (e.g. absorption intensity and bandwidth) of single gas components have been well established (e.g. high-resolution transmission molecular absorption, HITRAN, database published by Rothman et al., 2005); however, only limited studies assessed influences of dynamic temperature on gas quantification using OP-FTIR. Horrocks et al. (2001) and Smith et al. (2011) used OP-FTIR spectrometer and a gas cell purged with dry samples to measure the temperature-related error in SO$_2$, CO, CO$_2$, and CH$_4$ quantification, but not for N$_2$O.

To the best of our knowledge, there has never been a study of exploring the influence of changing temperature and path length on N$_2$O quantification under interferences of water vapour in spectra. To test the quality of OP-FTIR methods for multiple gas quantification is challenging due to interferences and the lack of proper measurement benchmarks. Also, the influence of water vapour is confounded by temperature and path length effects. In this study, the influence of water vapour, temperature, path length, and wind speed on N$_2$O and CO$_2$ quantification are examined using lab and field-based (OP-FTIR) methods.

2 Materials and experimental methods

2.1 Lab FTIR experiment

The lab FTIR experiment was used to assess the effects of water vapour and air temperature on N$_2$O quantification from spectra. An FTIR spectrometer equipped with a variable-path length gas cell was used to acquire reference spectra of N$_2$O, water vapour, and N$_2$O plus water vapour mixtures (i.e. wet N$_2$O) as shown in Fig. 1a.

2.1.1 Instrumentation setup

The lab FTIR spectrometer (Nexus 670, Thermo Electron Corporation, Madison, WI) was equipped with a KBr beam splitter, and a high-D* MCT detector was used to analyse gas samples using a multi-pass gas cell (White cell) (model MARS-8L/40L, Gemini Scientific Instruments, CA) with an optical path length of 33 m. Spectra range of 700.0-4000.0 cm$^{-1}$, optical velocity of 0.6 cm s$^{-1}$, and a resolution of 0.5 cm$^{-1}$ were selected for spectra acquisition. Each spectrum was acquired by co-adding 64 IFGs using the OMNIC software package (Thermo Fisher Scientific, Inc.) A triangular function was applied to apodize an IFG for converting an SB spectrum using the Fourier transform. A sampled SB spectrum contained the visible information of gases of interest and interferences. A background SB spectrum was collected from pure N$_2$ and used to convert a sampled SB spectrum to an absorbance spectrum. The temperature of the gas cell was controlled and monitored (model 689-0005, Barnant Co., Barrington, IL). The spectrometer was purged with dry air (−20 °C dew point) from a zero air generator (model 701H, Teledyne, Thousand Oaks, CA). Gas samples were produced using a diluter (series 4040, Envirionics Inc, Tolland, CT). The mole fraction (ppbv) of N$_2$O was diluted with ultra-pure N$_2$ gas. Water vapour concentrations (ppmv) were controlled by a
Nafion tube (Perma Pure, Lakewood, NJ) enclosed within a sealed container containing saturated water vapour. The concentration of saturated water vapour in the container was adjusted by controlling temperature of the water bath. Wet N\textsubscript{2}O gas samples were produced by passing dry N\textsubscript{2}O from the diluter through Nafion tube with a constant flow rate (4 L·min\textsuperscript{-1}). Gas samples were continuously introduced into the White cell with a constant pressure close to the room ambient pressure. Humidity and temperature probes (model HMT330, Vaisala Oyj, Helsinki, Finland) were used to monitor the humidity and temperature of the introduced gas samples. In this study, N\textsubscript{2}O (dry and wet) was diluted from 30.0 ppmv (N\textsubscript{2}O in N\textsubscript{2}) to 310.0, 400.0, 500.0, 600.0, and 700.0 ppbv with various water vapour contents (a relative humidity of 20 \%, 40 \%, 60 \%, and 80 \% at 30 °C).

### 2.1.2 Data collections and gas quantification

Spectra were collected when the N\textsubscript{2}O and water vapour concentrations and temperature of the introduced samples were steady. A total of nine single-gas absorption spectra, five dry N\textsubscript{2}O spectra with concentrations of 310.0-700.0 ppbv and four water vapour spectra with concentrations of 7000-28 000 ppmv were used to train CLS models (CSL-1 shown in Table 1). A total of sixty mixed-gas (wet N\textsubscript{2}O) spectra were used to build PLS models (see the table S2a in the supplement). Spectral windows (Table 2) and linear baseline correction were applied in CLS-1 and PLS models to calculate concentrations of N\textsubscript{2}O and water vapour from the mixed-gas spectra (the validation/wet sample spectra shown in Table 1) using TQ Analyst software Version 8.0 (Thermo Fisher Scientific, Inc.) The previous study showed that the integrated window of 2188.7-2204.1 and 2215.8-2223.7 cm\textsuperscript{-1} (W\textsubscript{N3}) is less sensitive to changing environment (e.g. water vapour = 5000-20 000 ppmv; ambient temperature = 10-35 °C) and provides higher accuracy than the window of 2170.0-2223.7 cm\textsuperscript{-1} (W\textsubscript{N1}) for N\textsubscript{2}O quantification using either CLS or PLS models (Lin et al., 2019). These two windows (W\textsubscript{N1} and W\textsubscript{N3}) were used to calculate N\textsubscript{2}O concentrations and examine the effect of water vapour, temperature, and path length on N\textsubscript{2}O calculations in this study. In PLS models, optimum factors were determined by cross validation and justified by the prediction of residual error sum of squares (PRESS) function to avoid over-fitting issues (see table S2b in the Supplement). Since it was difficult to isolate the effects of the temperature and humidity on quantitative biases from the field experiment, the validation spectra with the fixed concentrations of N\textsubscript{2}O and water vapour (310 ppbv N\textsubscript{2}O mixed with 21 500 ppmv water vapour) collected at 30, 35, 40 °C from the lab-FTIR to evaluate the sensitivity of CLS and PLS to temperature.

### 2.2 Open-path FTIR experiment

The OP-FTIR experiment was used to assess the effects of water vapour, air temperature, path lengths, and wind speed on the quantification of N\textsubscript{2}O and CO\textsubscript{2} from field spectra. The field instrumentation and configuration were shown in Fig. 1b.

#### 2.2.1 Site description and instrumentation setup

The field study was conducted at Purdue University Agronomy Center for Research and Education of West Lafayette, Indiana, the United States (86°56´ W, 40°49´ N). The predominant soil series at the study site was a Drummer silty clay loam (fine-
silty, mixed, mesic Typic Endoaquoll). The bulk density of topsoil (0-10 cm) and organic matter (0-20 cm) was measured 1.6 g·cm⁻³ and 3.8 %, respectively. The experimental site (Purdue Field Trace Gas Flux Facility) was between two maize cropping fields (~10 m apart) with anhydrous ammonia applications. A monostatic OP-FTIR (IR source, interferometer, transmitting and receiving telescope, and detector combined in one instrument) spectrometer (model 2501-C, MIDAC Corporation, Irvine, CA) was used to collect field-IR spectra. A retroreflector with 26 corner-cubes (cube-length of 76 mm) was mounted on a tripod 150 m from the spectrometer corresponding to an optical path of 300 m. The experiment of varying path lengths was conducted using optical path lengths of 100, 200, and 300 m.

### 2.2.2 Data collections and gas quantification

The same sample collection parameters were used to collect both OP-FTIR and lab-FTIR spectra. OP-FTIR spectra were collected using the AutoQuant Pro4.0 software package (MIDAC Corporation, Irvine, CA). Each field spectrum was collected by co-adding 64 IFGs and a resolution of 0.5 cm⁻¹. Triangular apodization, Mertz phase correction, and no zero-filling were applied to convert an IFG into an SB spectrum. A stray-light SB spectrum was collected by pointing the spectrometer telescope away from the retroreflector and subtracted from sampled SB spectra for stray-light correction. Quality control and assurance procedures (Russwurm, 1999; ASTM, 2013; Russwurm and Childers, 1999; Childers et al., 2001b; Shao et al., 2007) were used to evaluate spectra qualities and the influence of wind-induced vibrations. Gas concentrations derived from the OP-FTIR spectra were also calculated by CLS and PLS models. Training absorption spectra used in the CLS model were generated from the HITRAN database using E-trans (Ontar Corporation North Andover, MA). Briefly, high-resolution spectral lines of N₂O, CO₂, and water vapour output from E-trans were interpolated to generate spectra ranging from 500 cm⁻¹ to 4000 cm⁻¹ and convolved with a triangular apodization function. The convolved spectral lines were used to generate the reference spectra with the identical resolution and data point density matching the field spectra using Grams/32 (Childers et al., 2001). The HITRAN reference spectra were generated at the pressure of 760 torr and temperature of 30 °C.

The stray-light corrected field SB spectrum was converted to absorbance spectra by the synthetic SB background (syn-bkg) spectra using the IMACC Quantify package (Industrial Monitoring and Control Corp., Round Rock, TX). The syn-bkg was generated by selecting multiple points from the spectral interval of interest (i.e. six points within 2050.0-2500.0 cm⁻¹ for N₂O and CO₂) to fit the curvature of the sample SB spectrum using a polynomial function (Lin et al., 2019). Three spectral windows (Table 2) and the HITRAN references were used to build the CLS model (CLS-2 shown in Table 1) in the IMACC software. The third-degree polynomial function was used to correct the non-linear response of the CLS-calculated concentration to the actual concentration. More details regarding the IMACC quantification package were described in the IMACC user manual attached as the supplementary material. The PLS models were built using lab-FTIR measurements and only used for estimating N₂O concentrations. Since the molar fraction changes with changing air density which is the function of temperature and pressure, the measured temperature and pressure in the gas cell and experimental field were imported in the quantitative software to adjust the model-calculated concentrations.
Ambient temperature, relative humidity, and barometric pressure in the field were measured using an HMP45C probe (Vaisala Oyj, Helsinki, Finland) and a pressure sensor (278, Setra, Inc., Boxborough, MA) at 1.5 m above ground level (a.g.l.). The mean wind speed was measured by a 3-D sonic anemometer (model 81000, RM Young Inc., Traverse City, MI) mounted at 2.5 m a.g.l. and recorded at 16 Hz. A 50-m synthetic open path gas sampling system (S-OPS) (Heber et al., 2006) was used to collect gas samples along the OP-FTIR optical path to analyse the path-averaged concentrations of N₂O and CO₂ using a difference frequency generation mid-IR (DFG-IR) laser-based N₂O gas analyser (IRIS 4600, Thermo Fisher Scientific Inc., Waltham, MA) and a non-dispersive IR (ND-IR) CO₂ gas analyser (LI-840, LI-COR Inc., Lincoln, NE), respectively. The N₂O and CO₂ analysers provided a high precision for N₂O (< 0.15 ppbv, 1σ) and CO₂ (< 1.0 ppmv, 1σ) measurements, and both analysers were calibrated using the certified standard gas every four hours to insure the instrumentation stability and the accuracy. The path-averaged concentrations of N₂O and CO₂ were measured from both S-OPS and OP-FTIR simultaneously, and the S-OPS-measured concentrations were used as benchmarks to examine the accuracy and the sensitivity of OP-FTIR on gas quantification (Lin et al., 2019). The surface layer of air tends to become homogeneous in a well mixing condition (i.e. wind speed > 1.7 m s⁻¹ shown in Lin et al., 2019), and the well-mixed atmospheric condition can minimize spatial variations in the path-averaged concentrations from different measurement units (i.e., 50-m S-OPS vs. 150-m OP-FTIR).

2.2.3 Path lengths experiment

A variable path length between an OP-FTIR spectrometer and a retroreflector resulted in different path-integrated concentrations and the depth of gas absorbance in SB spectra. The complexities of N₂O absorption features within the 2170.0-2224.0 cm⁻¹ range required high spectral resolution (Fig. 3). For N₂O, the increased absorbance, resulting from a longer path length, likely improves its quantitative sensitivity and accuracy. Spectra were collected from physical lengths of 50, 100, and 150 m (Fig. 1b) using the same parameters. During the measurement (14:30-18:30, local time (LT) on 6 May 2016), background concentrations of N₂O (349.0 ± 0.5 ppbv) and CO₂ (400.0 ± 4.4 ppmv), ambient temperature and humidity (the relative humidity of 35.0 % at 25 °C) remained nearly constant. The spectra acquired from different path lengths were analysed by CLS-2 models for N₂O and CO₂ concentrations and by PLS models only for N₂O.

2.4 Quantitative accuracy

Quantitative accuracy/bias was determined by the relative error of the path-averaged concentrations between FTIR ($x_\text{f}$) and the true gas concentrations ($x_\text{t}$) of either the introduced gas (Lab) or the S-OPS measurements (Field), following Eq. (1):

$$\text{Bias} = [(x_\text{f} - x_\text{t})/x_\text{t}] \times 100 \%$$

(1)
3 Results and discussion

Quantification of target gas concentrations from either lab- or field-based FTIR spectra requires knowledge of the optimum spectral window (the spectral region used for quantification). In general, broadening the spectral windows will contain more spectral features that can be used for quantification. At the same time, however, broader windows will also contain more contributions from interfering constituents (e.g. water vapour). The optimum window would have clean spectral features of the target species with minimal spectral interference from other gases. For N₂O quantification, our previous work showed that the optimum window was to integrate two intervals of 2215.8-2223.7 and 2188.5-2204.1 cm⁻¹ (W₃ shown in Lin et al., 2019). In this paper, two windows (W₁ and W₃) and models (CLS and PLS) from the previous study were used to predict N₂O concentrations.

3.1 Lab FTIR experiment

3.1.1 Water vapour effect

Water vapour interfered with spectral windows and resulted in underestimations of N₂O concentrations using CLS models; increased water vapour increased the bias (Fig. 2). The PLS model provided more accurate predictions for N₂O than the CLS. The CLS accurately predicted gas concentrations only when the water vapour was absent or limited in spectra (Hong and Cho, 2003; Esler et al., 2000; Shao et al., 2010; Smith et al., 2011). In open-path measurements, CLS was often observed to underestimate gas concentrations, as reported by Childers et al. (2002), Briz et al. (2007), Shao et al. (2010), and Lin et al. (2019). Absorbance spectra of dry and wet N₂O showed that the interference of water vapour likely compromised the intensity of the N₂O P-branch absorbance (Fig. 3). Ideally, the absorbance/intensity of the 310 ppbv N₂O should be identical in either dry or wet conditions. The wet spectrum (red solid line), however, showed lower N₂O intensity than the dry spectrum (black solid line) after the baseline correction (Fig. 3). The reduced N₂O absorbance in wet samples resulted in underestimations of N₂O concentrations using CLS models created based on references of dry N₂O samples and water vapour. PLS models, created by wet N₂O references, showed improved accuracy in wet samples but overestimated N₂O in dry samples (Fig. 2). It is still unclear how water vapour interfered with gas quantification. The N₂O underestimation (based on the CLS model predictions) resulting from the attenuated absorbance was hypothesized due to the inadequate spectral resolution. High resolution is required to resolve rotation-vibrational gas spectral features (e.g. full-width at half height ~ 0.2 cm⁻¹) to avoid spectral nonlinearity to concentrations (ASTM, 2013; Griffiths and de Haseth, 2007; Russwurm and Phillip, 1999; Muller et al., 1999). Absorption features of N₂O were strongly overlapped by water vapour within 2170.0-2224.0 cm⁻¹. In order to resolve absorbance spectra of multiple gases and spectral overlaps, spectral resolution higher than 0.2 cm⁻¹ would be suggested. Increased optimal resolution, however, is a trade-off for the ratio of signal to noise which is along with detection limits as well as quantitative precision.
3.1.2 Temperature effect

The temperature-sensitivity of gas-phase FTIR spectra results in non-linearity of absorbance to temperature. Bias will be introduced if there is a temperature difference between reference and sample spectra (Russwurm and Phillip, 1999; Smith et al., 2011). The effect of this delta temperature on N₂O quantification is shown in Fig. 4a. Spectra of wet N₂O (310 ppbv N₂O blending with 21 000 ppmv water vapour) were collected at 30 °C, 35 °C, and 40 °C. Reference spectra of dry N₂O, water vapour, and wet N₂O were acquired at 30 °C and used to calculate N₂O concentrations from spectra collected at temperatures of 35 °C and 40 °C. The difference in temperature led to biases in N₂O concentrations (Fig. 4a). Smith et al. (2011) calculated concentrations of CO₂, CH₄, and CO using the MALT (Griffith, 1996) and showed that temperature-related error was approximately 3.0 % when the delta temperature was within 10 °C. As mentioned, water vapour present in spectra resulted in underestimations of N₂O by CLS models (Fig. 2), and this bias further increased with increasing the delta temperature. For instance, the bias increased from -3.0 to -5.0 % with increasing temperature from 30 to 40 °C using CLS models (Fig. 4a). Sources to this bias appeared to include 1) temperature-broadening of N₂O and 2) temperature-induced interference of water vapour (i.e. greater interference at increased temperature). Increased temperature tends to broaden the width of absorption lines in the rotation-vibration N₂O spectra and results in underestimations of gas concentrations. The increased strength of water vapour led to more interference in spectral analyses and great biases (Fig. 4b). Even though it is challenging to avoid interferences from water vapour, a proper window selection (e.g. W₅1 vs. W₅3) can mitigate the effect of water vapour on gas quantification. PLS methods also showed more accurate estimations and less sensitivity to temperature than CLS models (Fig. 4a).

3.2 Open-path FTIR experiment

3.2.1 Water vapour effect

In fields, water vapour ranged from 5000 to 20 000 ppmv during 9-19 June 2014. The increased water vapour showed the increased N₂O biases using CLS. In the lab experiment, water vapour increasing from 5000 to 20 000 ppmv at 30 °C only showed consistent underestimations of N₂O by approximately 3.0 % using CLS (Fig. 2). The positive relationship between water vapour content and the N₂O bias (Fig. 5a) was because water vapour was confounded by temperature. The increased temperature (i.e. 10-35 °C in fields) tends to increase humidity in the air. This negligible correlation (R² = 0.20 shown in Fig. 5a) became insignificant when the calculated biases were categorized by temperature (e.g. R² = 0.01 at the interval of 25-30 °C, data not shown). CO₂ measured by CLS was less sensitive to changing water vapour content (R² = 0.05) than N₂O (Fig. 5b) in field measurements, presumably due to the less water vapour absorption features in 2075.5-2084.0 cm⁻¹ than 2170.0-2224.0 cm⁻¹ (Lin et al., 2019). For PLS calculations, N₂O biases became consistent but slightly increased with increasing water vapour (Fig. 5c).
3.2.2 Temperature effect

Increased air temperature increased both N₂O and CO₂ biases from CLS models (Fig. 5d and 5e). A strong correlation of air temperature to N₂O biases (R² = 0.86) showed the N₂O quantification was more sensitive to temperature effects than CO₂ (R² = 0.39). The lab experiment (Fig. 4) showed that CLS underestimated N₂O by 3.0 % in wet air for samples with a low delta-temperature. N₂O calculations from OP-FTIR spectra, however, were underestimated by 12.0 % (approximately 36 ppbv less than the true value) even if the HITRAN reference and sample spectra were collected at the same temperature (i.e. 30 °C). The excess bias (12.0 % minus 3.0 %) likely resulted from interferences from CO and CO₂ in 2170.0-2224.0 cm⁻¹ and inherent uncertainties in line intensities and bandwidths of gas absorbance from HITRAN database (Rothman et al., 2005). The CLS-calculated CO₂ values were less influenced by temperature than N₂O (Fig. 5e), attributed to the reduced complexity of CO₂ absorption features in the 2075.0-2085.0 cm⁻¹ region, and less interference of water vapour within this region (Lin et al., 2019). Since the temperature-dependent absorption lines vary with species and wavelengths, the resolution parameter and gas quantification react differently to a changing environmental temperature (Smith et al., 2011; Griffith et al., 2012).

The PLS model were not only less affected by temperature (R² = 0.05) but provided better accuracy for N₂O estimations (Fig. 5f). For instance, N₂O bias was reduced from -12.0 % (the CLS-calculated N₂O shown in Fig. 5d) to 2.2 % using the PLS model at 30 °C (Fig. 5f). This slightly overestimated N₂O concentration (i.e. 2.2±0.8 % shown in Fig. 5f) was possibly due to 1) the limited information regarding other sources of environmental variables (e.g. wind-driven variables such as the homogeneity of the mixing air) in the PLS model, and 2) different background gases between training and field spectra. The pure N₂ was used as a buffer gas to generate reference mixtures (i.e. N₂O mixed with water vapour) to train the PLS model. For field spectra, however, only approximately 78 % N₂ is present in the spectra. Since the line-broadening coefficient is sensitive to the background compositions (Lacome et al., 1984), the discrepancy in backgrounds (i.e. N₂ vs. air) likely leads to errors in gas quantification. Furthermore, the linewidth of absorption bands is influenced by both temperature and pressure broadening effects (Loos et al., 2015). Variations in pressure between the reference and sample spectra also reduce the quantitative accuracy (Smith et al., 2011). In the control environment experiment, the pressure effect on bandwidth (or the associated error) was minimized due to the nearly constant pressure in the gas cell. The ambient air pressure was measured from 981.0 to 996.0 hPa every thirty minutes during the measurement period. Compared with the wide ranges of temperature and water vapour content, the effect of this small variations in the pressure (989.3±3.2 hPa, n = 355) on gas quantification was not examined, and the measured pressure was only used to adjust the model-calculated concentrations in this study.

3.2.3 Path length effect

OP-FTIR spectra containing N₂O concentrations of 349.0 ± 0.5 ppbv and CO₂ concentrations of 400.0 ± 4.4 ppmv were collected from optical lengths of 100, 200, and 300 m. As path lengths decreased, both N₂O and CO₂ concentrations were underestimated (Fig. 6a and 6b). For N₂O, CLS-derived concentrations were more sensitive to changing path lengths than PLS
(Fig. 6a). The Beer-Lambert law should show a linear response of absorbance to the path-integrated concentration. Nevertheless, the path-averaged absorbance of N₂O and CO₂ (i.e. \( Absorbance_{Path\ length(m)} \)) did not conform to the Beer-Lambert law even though background concentrations of N₂O and CO₂ were consistent (Fig. 6c and 6d), showing that there was a non-linear response of OP-FTIR spectra to the path-integrated concentrations. Several reasons may have caused non-linearity issues, such as detector saturation, spectral resolution, and apodization (ASTM, 2013; Russwurm and Childers, 1999; Griffiths and de Haseth, 2007). Detector saturation at short distances was avoided in this study by examinations of the IFG centre burst and SB spectra. For instance, the spectra were excluded if either the maximum or minimum signal of the centre burst exceeded the detector A/DC capacity (~ 2.5 V). Also, the elevated baseline below the detector cut-off, usually 600.0 cm⁻¹, in the SB spectrum was used as an indicator to inform the detector saturation (ASTM, 2013).

Thus, this short-path-derived bias (Fig. 6a and 6b) likely resulted from the inadequate spectral resolution. Short path lengths reduced the absorbance depth in an SB spectrum and sensitivity for quantification. Poorly resolved absorbance spectra could lead to a spectral non-linear response with different path-integrated concentrations (Zhu and Griffiths, 1994; Russwurm and Phillips, 1999). Also, N₂O quantification was more sensitive to path lengths than CO₂. With increasing optical path lengths from 100 m to 300 m, the accuracy of N₂O calculated from CLS models increased by approximately 20.0 % (N₂O biases reduced from −30.0 % to −10.0 % shown in Fig. 6a). For CO₂, the accuracy only increased 2.5 % (CO₂ biases reduced from −7.5 % to −5.0 % shown in Fig. 6b). The difference in sensitivity between gas quantification and path length was attributed to the complexity of absorbance spectra. N₂O absorption features in 2170.0-2224.0 cm⁻¹ were more complicated than CO₂ in 2075.0-2085.0 cm⁻¹; furthermore, more interfering gases (CO, CO₂, and water vapour in 2170.0-2224.0 cm⁻¹) interfered with N₂O quantification (Lin et al., 2019). A triangular apodization function applied in spectra results in a non-linear response (Russwurm and Phillips, 1999). The poorly-resolved spectra containing multiple gas species likely complicated the magnitude of the non-linearity led by apodization, which, however, was not evaluated by this study. The PLS methods reduced N₂O biases and showed less sensitivity to path length than CLS (Fig. 6a).

3.2.4 Wind speed effect

Nitrous oxide is predominately produced via soil microbial activities (nitrification and denitrification) and CO₂ is from respirations from soil microbes and vegetation (Mosier et al., 1996). As a result of soil and crop heterogeneities, multiple sources, and intermittent fluxes of N₂O and CO₂ from soil and/or canopy result in inhomogeneous gas concentrations in the atmosphere under low winds. Since the path length of the S-OPS (50-m) was different from the OP-FTIR (the physical length of 150 m), the gas uniformity across the 150 m influenced their path-averaged concentrations. The difference in the path-averaged N₂O and CO₂ concentrations between the S-OPS and OP-FTIR was used to calculate quantification bias (Fig. 7). Variabilities of N₂O and CO₂ biases were small but increased when the wind speed was less than 2 m s⁻¹ (Fig. 7c and Table S3 in the Supplement). This increased variability inferred the poorly-mixed air (< 2 m s⁻¹). Thus, decreasing wind speed and
turbulent mixing tended to increase gas concentration differences between the S-OPS and OP-FTIR. During the low wind environment, CO₂ bias showed higher variability than N₂O presumably due to a greater environmental variation in CO₂ concentrations than N₂O (Lin et al., 2019). For instance, CO₂ can be produced from both soil and canopy respirations, and plant uptake via photosynthesis. N₂O was predominately produced from soil nitrification and denitrification. The CO₂ concentrations and their spatial distribution in the air were influenced by the variabilities of both soil properties and crop species (different sources). Thus, CO₂ concentrations in the atmosphere tended to have higher variabilities than N₂O and become highly heterogeneous if the air was poorly mixed in the low wind condition. Low wind conditions likely occurred during the night period. During 9-19th 2014, a total of 259 and 130 data point (30-min averages N₂O) was collected from the daytime (06:00-20:00, LT) and nighttime (20:00-06:00, LT) measurements, respectively. The low wind conditions can occur during both day and night even though a low wind condition is more common during the night. In this study, 22% (57 out of 259) of all daytime measurements and 36% (47 out of 130) of all nighttime measurements were collected from low winds (< 2 m s⁻¹) (Fig. 7a).

4 Conclusion and recommendations

In this study, we have evaluated the effects of water vapour, temperature, path length, and wind speed on open-path FTIR measurements of N₂O and CO₂ quantified by CLS and PLS models. Water vapour in spectra underestimated N₂O concentrations by 3.0 % (lab experiment) and 12.0 % (field experiment) at 30 °C using CLS models. PLS models improved the accuracy of N₂O quantification (lab bias = −0.6 ± 0.4 % and field bias = 2.0 ± 0.8 %). Differences in temperature between reference and sample spectra led to errors in gas quantification. Increased air temperature significantly increased quantification bias using CLS models. For wet N₂O, 10 °C difference introduced 1.9 % (Lab) and 9.1 % (Field) more biases in gas concentrations. PLS models were less sensitive to temperature. Short path lengths reduced the sensitivity and accuracy for gas quantification, and CLS models were more sensitive to changing path lengths than PLS. These short-path-led biases were presumably due to the inadequate spectral resolution. CO₂ quantification using CLS model was less influenced by environmental variables than N₂O likely due to the less complex absorption features. The wind affected the mixings of gases and the low wind speed (< 2 m s⁻¹) led uncertainties in the path-averaged concentrations.

The partial least squares model generally provided more accurate measurements than CLS if the gas of interest is strongly interfered by water vapour (or other interferences) (e.g. strong overlap of water vapour absorbance features or broad spectral windows). PLS is also less sensitive to environmental variables than CLS. For OP-FTIR measurements, the CLS-calculated concentrations need to be verified carefully for quality assurance and to avoid substantial underestimations. Path lengths must be adequate, which can be checked by conducting a path length experiment before measurements. For the users interested in multi-source measurements, we suggested avoiding a great difference in path lengths if CLS models are used for gas quantification. High spectral resolution (< 0.5 cm⁻¹) is recommended to resolve complex spectral features of either gas of
interest or interferences. A high resolution also introduces more noises and increasing the scan time is suggested to increase the signal-to-noise ratio (Griffiths and de Haseth, 2007).

**Author contributions.** CHL, CTJ, RHG, and AJH designed the lab- and field-FTIR measurement experiments. CHL and RHG conducted the field experiment. CHL conducted the lab experiment, spectral and data analyses, and prepared the manuscript with contributions from CTJ, RHG, and AJH.

**Competing interests.** The authors declare that they have no conflict of interest.

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**References**

ASTM International.: ASTM E1865-97(2013) Standard Guide for Open-Path Fourier Transform Infrared (OP/FT-IR) Monitoring of Gases and Vapors in Air, West Conshohocken, PA, https://doi.org/10.1520/E1865-97R13, 2013a.

ASTM International.: ASTM E1982-98(2013) Standard Practice for Open-Path Fourier Transform Infrared (OP/FT-IR) Monitoring of Gases and Vapors in Air, West Conshohocken, PA, https://doi.org/10.1520/E1982-98R13, 2013b.

Baldocchi, D. D.: Assessing the eddy covariance technique for evaluating carbon dioxide exchange rates of ecosystems: past, present and future, Global Change Biology, 9, 479-492, https://doi.org/10.1046/j.1365-2486.2003.00629.x, 2003.

Bartoli, F., Allen, R., Esterowitz, L., and Kruer, M.: Auger-limited carrier lifetimes in HgCdTe at high excess carrier concentrations, J. Appl. Phys., 45, 2150-2154, https://doi.org/10.1063/1.1663561, 1974.

Bacsik, Z., Komlosi, V., Ollar, T., and Mink, J.: Comparison of open path and extractive long-path FTIR techniques in detection of air pollutants, Applied Spectroscopy Reviews, 41, 77-97, https://doi.org/10.1080/05704920500385494, 2006.

Briz, S., de Castro, A. J., Díez, S., López, F., and Schäfer, K.: Remote sensing by open-path FTIR spectroscopy. Comparison of different analysis techniques applied to ozone and carbon monoxide detection, J. Quant. Spectrosc. Radiat. Transf., 103, 314-330, https://doi.org/10.1016/j.jqsrt.2006.02.058, 2007.

Bjorneberg, D. L., Leytem, A. B., Westermann, D. T., Griffiths, P. R., Shao, L., and Pollard, M. J.: Measurement of atmospheric ammonia, methane, and nitrous oxide at a concentrated dairy production facility in southern Idaho using open-path FTIR spectrometry, Trans. ASABE, 52, 1749-1756, 2009.
Chase, D. B.: Nonlinear detector response in FT-IR, Appl. Spectrosc., 38, 491-494, https://doi.org/10.1366/0003702844555296, 1984.

Childers, J. W., Thompson, E. L., Harris, D. B., Kirchgessner, D. A., Clayton, M., Natschke, D. F., and Phillips, W. J.: Multi-pollutant concentration measurements around a concentrated swine production facility using open-path FTIR spectrometry, Atmos. Environ., 35, 1923-1936, https://doi.org/10.1016/s1352-2310(00)00545-8, 2001a.

Childers, J. W., Thompson, E. L., Harris, D. B., Kirchgessner, D. A., Clayton, M., Natschke, D. F., and Phillips, W. J.: Application of standardized quality control procedures to open-path fourier transform infrared data collected at a concentrated swine production facility, Environ. Sci. Technol., 35, 1859-1866, https://doi.org/10.1021/es001744f, 2001b.

Childers, J. W., Phillips, W. J., Thompson, E. L., Harris, D. B., Kirchgessner, D. A., Natschke, D. F., and Clayton, M.: Comparison of an Innovative Nonlinear Algorithm to Classical Least-Squares for Analyzing Open-Path Fourier Transform Infrared Spectra Collected at a Concentrated Swine Production Facility, Appl. Spectrosc., 56, 325-336, https://doi.org/10.1366/000370201954917, 2002.

de Castro, A. J., Lerma, A. M., López, F., Guijarro, M., Díez, C., Hernando, C., and Madrigal, J.: Open-path Fourier transform infrared spectrometry characterization of low temperature combustion gases in biomass fuels, Infrared Phys. Technol., 51, 21-30, https://doi.org/10.1016/j.infrared.2006.11.005, 2007.

Du, Y. P., Liang, Y. Z., Jiang, J. H., Berry, R. J., and Ozaki, Y.: Spectral regions selection to improve prediction ability of PLS models by changeable size moving window partial least squares and searching combination moving window partial least squares, Analytica Chimica Acta, 501, 183-191, https://doi.org/10.1016/j.aca.2003.09.041, 2004.

Esler, M. B., Griffith, D. W. T., Wilson, S. R., and Steele, L. P.: Precision trace gas analysis by FT-IR spectroscopy. 1. Simultaneous analysis of CO₂, CH₄, N₂O, and CO in air, Analytical Chemistry, 72, 206-215, https://doi.org/10.1021/ac990625, 2000.

Flesch, T. K., Baron, V. S., Wilson, J. D., Griffith, D. W. T., Basarab, J. A., and Carlson, P. J.: Agricultural gas emissions during the spring thaw: Applying a new measurement technique, Agric. For. Meteorol., 221, 111-121, https://doi.org/10.1016/j.agrformet.2016.02.010, 2016.

Denmead, O. T.: Approaches to measuring fluxes of methane and nitrous oxide between landscapes and the atmosphere, Plant Soil, 309, 5-24, https://doi.org/10.1007/s11104-008-9599-z, 2008.

Griffith, D. W.: Synthetic calibration and quantitative analysis of gas-phase FT-IR spectra, Appl. Spectrosc., 50, 59-70, https://doi.org/10.1366/0003702963906627, 1996.

Griffith, D. W. T., and Jamie, I. M.: Fourier Transform Infrared Spectrometry in Atmospheric and Trace Gas Analysis, in: Encyclopedia of Analytical Chemistry, https://doi.org/10.1002/9780470027318.a0710, John Wiley & Sons, Hoboken, 2006.

Griffith, D. W. T., Deutscher, N. M., Caldow, C., Kettlewell, G., Riggenbach, M., and Hammer, S.: A Fourier transform infrared trace gas and isotope analyser for atmospheric applications, Atmospheric Measurement Techniques, 5, 2481-2498, https://doi.org/10.5194/amt-5-2481-2012, 2012.
Griffiths, P. R.: "Fourier Transform spectrometry at low resolution: how low can you go?", Proc. SPIE 2089, 9th International Conference on Fourier Transform Spectroscopy, https://doi.org/10.1117/12.166563, 1994.
Griffiths, P. R., and de Haseth, J. A.: Fourier transform infrared spectrometry, John Wiley & Sons, Hoboken, 2007.
Griffiths, P., Shao, L., and Leytem, A.: Completely automated open-path FT-IR spectrometry, Analytical and bioanalytical chemistry, 393, 45-50, https://doi.org/10.1007/s00216-008-2429-6, 2009.

Haaland, D. M.: Methods to Include Beer's Law Nonlinearities in Quantitative Spectral Analysis, in, edited by: McClure, G. L., ASTM International, West Conshohocken, PA, 78-94, https://doi.org/ 10.1520/STP23164S, 1987.
Haaland, D. M., and Easterling, R. G.: Improved Sensitivity of Infrared Spectroscopy by the Application of Least Squares Methods, Appl. Spectrosc., 34, 539-548, https://doi.org/10.1366/0003702804731258, 1980.

Haaland, D. M., and Thomas, E. V.: Partial least-squares methods for spectral analyses. 1. Relation to other quantitative calibration methods and the extraction of qualitative information, Anal. Chem., 60, 1193-1202, https://doi.org/10.1021/ac00162a020, 1988.
Hart, B. K., and Griffiths, P. R.: Use of partial least squares regression for the multivariate calibration of hazardous air pollutants in open-path FT-IR spectrometry, AIP Conference Proceedings, 430, 241-244, https://doi.org/10.1063/1.55811, 1998.
Hart, B. K., Berry, R. J., and Griffiths, P. R.: Effects of resolution, spectral window, and background on multivariate calibrations used for open-path Fourier transform infrared spectrometry, Field Analytical Chemistry and Technology, 3, 117-130, https://doi.org/10.1002/(sici)1520-6521(1999)3:2<117::aid-fact7>3.0.co;2-v, 1999.
Hart, B. K., Berry, R. J., and Griffiths, P. R.: Effects of resolution on quantification in open-path Fourier transform infrared spectrometry under conditions of low detector noise. 2. Partial least squares regression, Environmental Science & Technology, 34, 1346-1351, https://doi.org/10.1021/es990439v, 2000.
Hart, B. K., and Griffiths, P. R.: Effect of resolution on quantification in open-path Fourier transform infrared spectrometry under conditions of low detector noise. 1. Classical least squares regression, Environ. Sci. Technol., 34, 1337-1345, https://doi.org/10.1021/es9904383, 2000.

Hong, D. W., and Cho, S. Y.: Improved methods for performing multivariate analysis and deriving background spectra in atmospheric open-path FT-IR monitoring, Applied Spectroscopy, 57, 299-308, 10. https://doi.org/10.1366/00370203321558218, 2003.
Heber, A. J., Ni, J. Q., Lim, T. T., Tao, P. C., Schmidt, A. M., Koziel, J. A., Beasley, D. B., Hoff, S. J., Nicolai, R. E., Jacobson, L. D., and Zhang, Y. H.: Quality assured measurements of animal building emissions: Gas concentrations, J. Air Waste Manage. Assoc., 56, 1472-1483, https://doi.org/10.1080/10473289.2006.10465680, 2006.
Horrocks, L. A., Oppenheimer, C., Burton, M. R., Duffell, H. J., Davies, N. M., Martin, N. A., and Bell, W.: Open-path Fourier transform infrared spectroscopy of SO2: An empirical error budget analysis, with implications for volcano monitoring, J. Geophys. Res.-Atmos., 106, 27647-27659, https://doi.org/10.1029/2001jd000343, 2001.
IPCC: Intergovernmental Panel on Climate Change, Fourth Assessment Report-Working I Report: The physical science basis, Cambridge Univ. Press, Cambridge, 2007.

IPCC, Climate Change 2014: Mitigation of Climate Change. Contribution of Working Group III to the Fifth Assessment Report of the Intergovernmental Panel on Climate Change Cambridge University Press, Cambridge, United Kingdom and New York, NY, USA, <http://www.ipcc.ch/pdf/assessment-report/ar5/wg3/ipcc wg3 ar5_full pdf>, 2014.

Jiang, J. H., Berry, R. J., Siesler, H. W., and Ozaki, Y.: Wavelength interval selection in multicomponent spectral analysis by moving window partial least-squares regression with applications to mid-infrared and near-infrared spectroscopic data, Analytical Chemistry, 74, 3555-3565, https://doi.org/10.1021/ac011177u, 2002.

Kosterev, A., Wysocki, G., Bakhirkin, Y., So, S., Lewicki, R., Fraser, M., Tittel, F., and Curl, R. F.: Application of quantum cascade lasers to trace gas analysis, Appl. Phys. B-Lasers Opt., 90, 165-176, https://doi.org/10.1007/s00340-007-2846-9, 2008.

Lacome, N., Levy, A., and Guelachvili, G.: Fourier transform measurement of self-, N2-, and O2-broadening of N2O lines: temperature dependence of linewidths, Appl. Optics, 23, 425-435, https://doi.org/10.1364/AO.23.000425, 1984.

Lam, S. K., Suter, H., Mosier, A. R., and Chen, D.: Using nitrification inhibitors to mitigate agricultural N2O emission: a double-edged sword?, Global Change Biology, 23, 485-489, https://doi.org/10.1111/gcb.13338, 2017.

Lin, C. H., Grant, R. H., Heber, A. J., and Johnston, C. T.: Application of open-path Fourier transform infrared spectroscopy (OP-FTIR) to measure greenhouse gas concentrations from agricultural fields, Atmos. Meas. Tech., 12, 3403-3415, https://doi.org/10.5194/amt-12-3403-2019, 2019.

Loos, J., Birk, M., and Wagner, G.: Pressure broadening, shift, speed dependence and line mixing in the v(3) rovibrational band of N2O, Journal of Quantitative Spectroscopy & Radiative Transfer, 151, 300-309, https://doi.org/10.1016/j.jqsrt.2014.10.008, 2015.

Mosier, A. R., Duxbury, J. M., Freney, J. R., Heinemeyer, O., and Minami, K.: Nitrous oxide emissions from agricultural fields: Assessment, measurement and mitigation, Plant and Soil, 181, 95-108, https://doi.org/10.1007/bf00011296, 1996.

Muller, U., Heise, H. M., Mosebach, H., Gartner, A. G., and Hausler, T.: Improved strategies for quantitative evaluation of atmospheric FTIR spectra obtained in open-path monitoring, Field Anal. Chem. Technol., 3, 141-159, https://doi.org/10.1002/(sici)1520-6521(1999)3:3<141::aid-fact2>3.0.co;2-5, 1999.

Nelson, D. D., Zahniser, M. S., McManus, J. B., Shorter, J. H., Wormhoudt, J. C., and Kolb, C. E.: Recent improvements in atmospheric trace gas monitoring using mid-infrared tunable diode lasers, Proc. SPIE 2834, 148-159, https://doi.org/10.1117/12.255321, 1996.

Paton-Walsh, C., Smith, T. E. L., Young, E. L., Griffith, D. W. T., and Guerette, E. A.: New emission factors for Australian vegetation fires measured using open-path Fourier transform infrared spectroscopy - Part I: Methods and Australian temperate forest fires, Atmospheric Chemistry and Physics, 14, 11313-11333, https://doi.org/10.5194/acp-14-11313-2014, 2014.
Phillips, F. A., Naylor, T., Forehead, H., Griffith, D. W. T., Kirkwood, J., and Paton-Walsh, C.: Vehicle Ammonia Emissions Measured in An Urban Environment in Sydney, Australia, Using Open Path Fourier Transform Infra-Red Spectroscopy, Atmosphere, 10, https://doi.org/10.3390/atmos10040208, 2019.

Rothman, L. S., Jacquemart, D., Barbe, A., Benner, D. C., Birk, M., Brown, L. R., Carleer, M. R., Chackerian, C., Chance, K., Coudert, L. H., Dana, V., Devi, V. M., Flaud, J. M., Gamache, R. R., Goldman, A., Hartmann, J. M., Jucks, K. W., Maki, A. G., Mandin, J. Y., Massie, S. T., Orphal, J., Perrin, A., Rinsland, C. P., Smith, M. A. H., Tennyson, J., Tolchenov, R. N., Toth, R. A., Vander Auwera, J., Varanasi, P., and Wagner, G.: The HITRAN 2004 molecular spectroscopic database, J. Quant. Spectrosc. Radiat. Transf., 96, 139-204, https://doi.org/10.1016/j.jqsrt.2004.10.008, 2005.

Russwurm, G. M., and Childers, J. W.: FT-IR open-path monitoring guidance document (3rd edition), ManTech Environmental Technology, Inc., Research Triangle Park, NC (United States), 1999.

Russwurm, G. M., and Phillips, B.: Effects of a nonlinear response of the Fourier-transform infrared open-path instrument on the measurements of some atmospheric gases, Appl. Optics, 38, 6398-6407, https://doi.org/10.1364/ao.38.006398, 1999.

Russwurm, G. M.: Long-path open-path Fourier transform infrared monitoring of atmospheric gases: compendium method TO-16, in: Compendium of Methods for the Determination of Toxic Organic Compounds in Ambient Air, 2nd ed., Center for Environmental Research Information, Office of Research and Development, US Environmental Protection Agency, Cincinnati, OH, EPA/625/R-96/010b, 1999.

Russwurm, G. M. and Childers, J. W.: “Open-path Fourier transform infrared spectroscopy”. In Handbook of Vibrational Spectroscopy Edited by: Chalmers, J. M. and Griffiths, P. R. Vol. 2, 1750–1773. New York: Wiley, 2002.

Shao, L., Pollard, M. J., Griffiths, P. R., Westermann, D. T., and Bjorneberg, D. L.: Rejection criteria for open-path Fourier transform infrared spectrometry during continuous atmospheric monitoring, Vib. Spectrosc., 43, 78-85, https://doi.org/10.1016/j.vibspec.2006.06.016, 2007.

Shao, L., Griffiths, P. R., and Leytem, A. B.: Advances in Data Processing for Open-Path Fourier Transform Infrared Spectrometry of Greenhouse Gases, Anal. Chem., 82, 8027-8033, https://doi.org/10.1021/ac101711r, 2010.

Smith, T. E. L., Wooster, M. J., Tattaris, M., and Griffith, D. W. T.: Absolute accuracy and sensitivity analysis of OP-FTIR retrievals of CO₂, CH₄ and CO over concentrations representative of "clean air" and "polluted plumes", Atmos. Meas. Tech., 4, 97-116, https://doi.org/10.5194/amt-4-97-2011, 2011.

Zhu, C., and Griffiths, P. R.: Extending the Range of Beer's Law in FT-IR Spectrometry. Part I: Theoretical Study of Norton-Beer Apodization Functions, Appl. Spectrosc. 52, 1403-1408, https://doi.org/10.1366/003702981943022, 1998.
Table 1. Multivariate models (i.e. CLS and PLS) used for N\textsubscript{2}O and CO\textsubscript{2} quantification: 1) the CLS-1 model was built by five dry N\textsubscript{2}O spectra (310, 400, 500, 600, 700 ppbv) and four water vapour spectra (7 000, 15 000, 22 000, 28 000 ppmv) collected from the lab-FTIR, 2) CLS-2 model was built by twelve dry N\textsubscript{2}O, seventeen dry CO\textsubscript{2}, and sixteen water vapour spectra generated from the HITRAN database (see Table S1 in the supplement), and 3) the PLS model was built by a total of sixty wet N\textsubscript{2}O spectra collected from Lab-FTIR (see Table S2 in the supplement).

| Model | Training/Reference spectra | Source | Validation/Wet Sample spectra | Source |
|-------|-----------------------------|--------|-------------------------------|--------|
| CLS-1 | Dry N\textsubscript{2}O: 310-700 ppbv  
Water vapour: 7,000-28,000 ppmv | Lab-FTIR | N\textsubscript{2}O | Lab-FTIR |
| CLS-2 | Dry N\textsubscript{2}O, CO\textsubscript{2}, water vapour: see Table S1 | HITRAN | N\textsubscript{2}O, CO\textsubscript{2} | OP-FTIR |
| PLS   | Wet N\textsubscript{2}O: see Table S2 | Lab-FTIR | N\textsubscript{2}O | Lab- and OP-FTIR |
**Table 2.** Spectral windows for quantification of N$_2$O and CO$_2$.

| Gas   | Windows (cm$^{-1}$)                                                                 | Interferences       |
|-------|-----------------------------------------------------------------------------------|---------------------|
| N$_2$O | $W_N^1$: 2170.0-2223.7                                                            | H$_2$O, CO, CO$_2$  |
|       | $W_N^3$: 2188.7-2204.1 + 2215.8-2223.7                                            | H$_2$O, CO, CO$_2$  |
| CO$_2$ | $W_C^2$: 2075.5-2084.0                                                             | H$_2$O              |
Figure 1. Schematic of the instrumentation used to assess the effects of water vapour and temperature on gas quantification: (a) lab-FTIR with a multi-pass gas cell (optical path length of 33 m); (b) DFG-IR N₂O and ND-IR CO₂ analysers combined with synthetic open path gas sampling system (S-OPS) were used as benchmarks to assess quantification of N₂O and CO₂ from OP-FTIR.
**Figure 2.** Effects of water vapour on N₂O quantification: the lab-FTIR spectra of dry N₂O and N₂O/water vapour mixtures (310 ppbv N₂O at the relative humidity of 20%, 40%, 60%, and 80% at 30 °C) were used to calculate N₂O concentrations using CLS and PLS models and two spectral windows (WN1 and WN3).
Figure 3. The lab-FTIR spectra of dry N\textsubscript{2}O (310 ppbv), wet N\textsubscript{2}O (310 ppbv N\textsubscript{2}O plus 28 000 ppmv water vapour), and water vapour (28 000 ppmv) were acquired at 30 °C.
Figure 4. Effects of temperature on N₂O quantification: the lab FTIR spectra containing wet N₂O (310 ppbv N₂O plus 21 500 ppmv water vapour) were acquired at 30 °C, 35 °C, and 40 °C. Temperature affected (a) N₂O concentrations calculated by the CLS and PLS models and two spectral windows (Wₙ₁ and Wₙ₃) and (b) spectral differences in wet N₂O absorbance (310 ppbv N₂O plus 21 500 ppmv water vapour) among different temperature.
Figure 5. Effects of water vapour and temperature on N₂O and CO₂ concentrations calculated from the OP-FTIR spectra using CLS and PLS models and the optimum windows (W₃ for N₂O and W₂ for CO₂) during 9-19 June 2014. Assumed temperature and bias are a linear relationship.
Figure 6. Effects of path lengths on N$_2$O and CO$_2$ quantification: OP-FTIR spectra were acquired under backgrounds of N$_2$O (349.0 ± 0.5 ppbv) and CO$_2$ (400.0 ± 4.4 ppmv), and relative humidity of 35% at 25°C from optical path lengths of 100, 200, and 300 m for quantification of (a) N$_2$O using CLS and PLS and windows of W$_N^1$ and W$_N^3$, (b) CO$_2$ using a CLS model and the window of W$_C^2$. The path-averaged absorbance spectra of (c) N$_2$O and (d) CO$_2$ showed the inconsistency of absorbance spectra.
Figure 7. Effects of the wind speed on differences in the path-averaged concentrations between the OP-FTIR and S-OPS: (a) N₂O, (b) CO₂, and (c) variability of biases (Standard deviation, SD).