Decontamination of water co-polluted by copper, toluene and tetrahydrofuran using lauric acid

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Co-contamination by organic solvents (e.g., toluene and tetrahydrofuran) and metal ions (e.g., Cu2+) is common in industrial wastewater and in industrial sites. This manuscript describes the separation of THF from water in the absence of copper ions, as well as the treatment of water co-polluted with either THF and copper, or toluene and copper. Tetrahydrofuran (THF) and water are freely miscible in the absence of lauric acid. Lauric acid separates the two solvents, as demonstrated by proton nuclear magnetic resonance (1H NMR) and Attenuated Total Reflection-Fourier Transform Infrared Spectroscopy (ATR-FTIR). The purity of the water phase separated from 3:7 (v/v) THF:water mixtures using 1 M lauric acid is ≈87%v/v. Synchrotron small angle X-ray scattering (SAXS) indicates that lauric acid forms reverse micelles in THF, which swell in the presence of water (to host water in their interior) and ultimately lead to two free phases: 1) THF-rich and 2) water-rich. Deprotonated lauric acid (laurate ions) also induces the migration of Cu2+ ions in either THF (following separation from water) or in toluene (immiscible in water), enabling their removal from water. Laurate ions and copper ions likely interact through physical interactions (e.g., electrostatic interactions) rather than chemical bonds, as shown by ATR-FTIR. Inductively coupled plasma—optical emission spectrometry (ICP-OES) demonstrates up to 60% removal of Cu2+ ions from water co-polluted by CuSO4 or CuCl2 and toluene. While lauric acid emulsifies water and toluene in the absence of copper ions, copper salts destabilize emulsions. This is beneficial, to avoid that copper ions are re-entrained in the water phase alongside with toluene, following their migration in the toluene phase. The effect of copper ions on emulsion stability is explained based on the decreased interfacial activity and compressional rigidity of interfacial films, probed using a Langmuir trough. In wastewater treatment, lauric acid (a powder) can be mixed directly in the polluted water. In the context of groundwater remediation, lauric acid can be solubilized in canola oil to enable its injection to treat aquifers co-polluted by organic solvents and Cu2+. In this application, injectable filters obtained by injecting cationic hydroxyethylcellulose (HEC +) would impede the flow of toluene and copper ions partitioned in it, protecting downstream receptors. Co-contaminants can be subsequently extracted upstream of the filters (using pumping wells), to enable their simultaneous removal from aquifers.

Industrial activities release water-soluble, toxic heavy metals in groundwater, including lead, chromium, arsenic, zinc, cadmium, mercury and copper1–2. Copper is used in fertilizers and pesticide sprays, building materials, and agricultural and municipal waste, causing high copper concentrations in groundwater1. Hydrocarbons are also widely used in industrial processes and they are amongst the most common groundwater pollutants1. THF is a groundwater and industrial wastewater contaminant, because it used to produce pharmaceutical and pesticide intermediates4–5. Heavy metals, hydrocarbons and water-miscible organic solvents (e.g., dioxane or THF) are often present as co-contaminants on industrial sites6–11.

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Treatment of heavy metals includes electrokinetic remediation\textsuperscript{12–16}, removal using nanoparticles\textsuperscript{17} and soil flushing with additives which facilitate the solubilization and extraction of heavy metals through pump and treat\textsuperscript{1}. Pump and treat extracts pollutants using pumping wells, treats groundwater ex situ and finally injects it after treatment\textsuperscript{18}. As an example, ethylenediaminetetraacetic acid (EDTA) has been used to remediate copper in conjunction with pump and treat\textsuperscript{19}. Our previous study used sodium lauroyl lactylate (SLL) for the same purpose\textsuperscript{20}. These approaches do not enable the simultaneous removal of heavy metals and co-contaminants such as miscible solvents, of which THF is an example.

Common remedial approaches to target hydrocarbon contamination include chemical methods (e.g., chemical degradation using Fenton's reagents\textsuperscript{21,22}, oxidizing emulsifiers\textsuperscript{23}, slow release oxidizer rods\textsuperscript{24}, oxidizing nanoparticles\textsuperscript{25}, ferrate\textsuperscript{26} and ozone sparging\textsuperscript{27}), phytoremediation for shallow contaminants\textsuperscript{28}, microwaves heating\textsuperscript{29}, pump and treat\textsuperscript{30}, surfactant flushing\textsuperscript{11,32} and surfactant enhanced oxidation\textsuperscript{33}, bioremediation using bacteria\textsuperscript{34,35}, smouldering\textsuperscript{36} and thermal treatment\textsuperscript{37}. Electrokinetic remediation of hydrocarbons has also been proposed, for instance in conjunction with surfactants\textsuperscript{38,39}. Hydrocarbons can migrate during their remediation, posing risks to downstream receptors\textsuperscript{40–42}. This risk also exists for heavy metals.

Our previous research focused on immobilizing Cr(VI) during its reduction to less toxic Cr(III)\textsuperscript{43}, and barriers can mitigate hydrocarbon migration during treatment\textsuperscript{44}. Barriers can be non-reactive\textsuperscript{18} or reactive\textsuperscript{45–50}. In our previous studies focusing on hydrocarbons we also developed semi-permeable filters that could be injected for facile installation\textsuperscript{44}. These filters enabled water flow, while retaining hydrocarbons\textsuperscript{51}. While useful to prevent hydrocarbon migration, these filters could not simultaneously retain heavy metals.

Organic solvents miscible in water can be removed from wastewater using biological treatment\textsuperscript{52–54}, chemical methods (e.g., advanced oxidation processes\textsuperscript{55}) or pervaporation\textsuperscript{56,57}. Biological treatment requires that pollutant concentrations are below levels at which they are toxic to the bacteria. Advanced oxidation processes are useful but require potentially hazardous chemicals and pervaporation is energy intensive.

Moreover, separation between water and organic solvents can be achieved by enhancing the organic solvent hydrophobicity (i.e., strengthening the organic molecule network at the expense of water- solvent interactions). For instance, a study showed that temperature could be tuned to decrease the ionicity of the lidocaine-oleic acid phase in a mixture of water, lidocaine and oleic acid, thereby separating it into a water phase and a lidocaine-oleic acid phase\textsuperscript{58}. It would be advantageous to achieve separation without temperature cycles. Previous studies separated water from water-miscible solvents using salts\textsuperscript{59,60}, sugars\textsuperscript{61,62} and polyols\textsuperscript{63} and surfactants such as sodium lauroyl lactylate (SLL)\textsuperscript{64}. These compounds hydrogen bond (H-bonds) with water, thereby hindering water-solvent interactions\textsuperscript{64–69}. In these previous studies, copper ions were not separated from water alongside with THF.

Liquid–liquid extraction has been applied to other contaminants such as antibiotics, pesticides and hydrocarbons, to facilitate their detection through analytical methods\textsuperscript{70–72}, but the simultaneous separation of THF and metal ions (e.g., copper) was not previously achieved in the context of water purification. Here, we use for the first time lauric acid to simultaneously separate THF from water and induce copper ion migration in the THF phase, at ambient temperature (e.g., 20 °C). The proposed approach has potential applications to treat wastewater, surface water or groundwater, for instance following its extraction from polluted aquifers through pump and treat.

Moreover, we use injectable filters in conjunction with lauric acid to simultaneously target heavy metal contamination (e.g., Cu\textsuperscript{2+}) and hydrocarbons immiscible in water. Specifically, we inject lauric acid in canola oil in model polluted aquifers, to induce Cu\textsuperscript{2+} migration in toluene (used as model hydrocarbon), enabling its retention by injectable HEC+ filters. Toluene and Cu\textsuperscript{2+} could then be removed more easily from groundwater, e.g., by pumping them in filter proximity, upstream of the filters. Lauric acid\textsuperscript{73} and oleic acid\textsuperscript{74} were previously used to extract Cu\textsuperscript{2+} from water, but their use was not combined with injectable barriers. Also, our previous research showed that the benign surfactant SLL also induced Cu\textsuperscript{2+} partitioning into hydrophobic solvents such as canola oil, used as liquid sorbent in the context of water purification\textsuperscript{20}. However, in this case, the oil phase became solid and could not be pumped from the subsurface. Therefore, SLL was exclusively intended for ex situ water treatment following pump and treat\textsuperscript{90}.

**Materials and methods.**

**Materials.** CuCl\textsubscript{2}·2H\textsubscript{2}O and CuSO\textsubscript{4}·5H\textsubscript{2}O (ACS grade), lauric acid (ACS grade), toluene (HPLC grade), NaOH (pellets, ACS grade) and cationic hydroxyethyl cellulose (HEC+) were purchased from Sigma Aldrich (Canada). Canola oil (Selection brand) was purchased from a local market. THF (Caledon Laboratories Ltd, Canada) and dioxane (Fisher Scientific Canada, ACS grade) were purchased from Fisher Scientific (Canada). Deionised (DI) water was used in all experiments.

**Bottle tests.** Bottle tests were conducted in glass vials, to investigate separation between THF and water, using 30%, 50% and 70% THF relative to water (v/v) and lauric acid concentrations ranging from 0.2 to 1 M, at pH = 2, 6.5 and 13. Equation (1) was used to estimate the difference between the volume of water added to the mixture (\(V_{\text{water,used}}\)) and the volume of the water-rich bottom phase separated using lauric acid (\(V_{\text{water,measured}}\))

\[
\frac{(V_{\text{water,measured}} - V_{\text{water,used}})}{V_{\text{water,used}}} \times 100(\%)
\]  

(1)

In the “Results and discussion” section this difference is referred to as deviation.

Bottle tests were also conducted to assess the partitioning of Cu\textsuperscript{2+} ions between the water and oil phases, and to study emulsification between toluene and water. Bottle tests were conducted by mixing first 5 mL of 30 M aqueous CuCl\textsubscript{2} or CuSO\textsubscript{4} solutions, 1 mL of 0.23 M lauric in canola oil solutions and either 1 mL or 4 mL of toluene. Afterwards, 10 mM NaOH (relative to the water phase) were added, and vials were agitated by hand for 30 s.
before equilibrating them on a bench. Experiments were also conducted with pure water (without copper salts), using 1 mL of 0.23 M lauric in canola oil solutions and 1 mL of toluene, without NaOH or with 10 mM NaOH. Finally, additional bottle tests were conducted using 30 mM CuCl₂ or CuSO₄ solutions, toluene, and lauric acid alone, at pH = 13. The volume ratio of the aqueous phase to the toluene phase was four parts of water and one of toluene. All bottle tests were done in duplicate. Duplicates closely resembled each other.

### Static interfacial tension measurement.

Static interfacial tension measurements were conducted using a Sigma force tensiometer (Biolin Scientific, USA) and a platinum du Noüy ring (having a 2 cm diameter), using four types of samples, the composition of which is summarized in Table 1. The water and oil phases were mixed (1:1.2 water:oil, v/v) and allowed to separate in a separatory funnel for 24 h. They were then collected separately and carefully transferred in a glass cup with the aid of a pipette, to measure interfacial tension.

The surface tension provided in the “Results and discussion” section was estimated using (2):

\[
\gamma = \frac{F}{4\pi R^2} \tag{2}
\]

where \( F \) is the maximum force measured when pulling the du Noüy ring outside of the water phase (i.e., the force required to break the lamella), \( R \) is the average radius of the Du Noüy ring used and \( f \) is the Huh and Mason correction factor calculated as \( f = \frac{R}{r} \) (\( R \) = radius of the Du Noüy ring and \( r \) = radius of the wire). All measurements were conducted in triplicate, to estimate the low standard deviation reported in the “Results and discussion” section.

### Compression isotherms.

Compression isotherms were measured at the oil–water interface using a Kibron Microtrough G1 Langmuir–Blodgett trough (Kibron, Sweden), controlled using KBN LayerXPro software (Kibron, Sweden). The water and oil phases are as described for samples 3 and 4 used for static interfacial tension measurements (Table 1, “Static interfacial tension measurement”). Samples were prepared in a separatory funnel, agitated and allowed to separate overnight. The water and oil phases were then collected separately, and carefully transferred into the lower and upper compartment of the Langmuir trough, respectively (with the aid of a pipette). After transferring the top and bottom phases, interfacially active species were allowed to adsorb at the interface for 10 min, before being compressed from 16,500 to 1650 mm², using mobile barriers moving at a speed of 20 mm/min. After the first compression the barriers were rapidly expanded (at a speed of 140 mm/min) and films were immediately recompressed at 20 mm/min. The pressure was monitored using a Wilhelmy plate during each compression. Measurements were conducted in duplicate and measured curves closely resembled each other.

### ATR-FTIR.

ATR-FTIR measurements were conducted to investigate potential interactions between lauric acid and Cu²⁺ ions. In these measurements, the samples used were prepared as described in “Bottle tests”. Specifically, we mixed first 5 mL of 30 M aqueous CuCl₂ or CuSO₄ solutions, 1 mL of 0.23 M lauric acid in canola oil solutions and 1 mL of toluene. Afterwards, 10 mM NaOH (relative to the water phase) were added, and vials were agitated by hand for 30 s, before allowing to settle on the bench. Experiments were also conducted without copper salts, using 1 mL of 0.23 M lauric acid in canola oil solutions and 1 mL of toluene, with 10 mM NaOH. Only the oil phase was analyzed, after 24 h equilibration. Additionally, ATR-FTIR was used to analyze samples prepared using 3:7 THF:water mixtures (v/v), using either water or 0.03 M CuCl₂ or CuSO₄ aqueous solutions. Absorbance spectra were collected using an ATR-FTIR spectrometer (Thermoscientific Nicolet Summit FTIR spectrometer with an Everest ATR), with an accompanying IR solution software. Due to the volatility of THF, each spectrum is the average of only 10 scans, with a resolution of 4 cm⁻¹, in the wavenumber range of 400–4000 cm⁻¹. Measurements were repeated at least four times per sample to account for the low scan count.

The absorbance peak associated with the OH stretch (H-bonding) of 3:7 v/v THF:water mixtures (pH ≈ 6.5 and pH ≈ 3) separated with 1 M lauric acid was deconvolved using a coarse-grain approach developed by us, as described in our previous manuscripts. Briefly, normalized experimental data in the range 2500–4000 cm⁻¹ were fitted to a sum of two Gaussians by nonlinear regression using Graphpad Prism 9.2.0 (Eq. 3). This software uses the Levenberg and Marquardt algorithm. We excluded the C–H stretching intensities in the range 2650–3000 cm⁻¹ to improve the stability of the fit considerably and remove non-water OH contributions from the analysis.

\[
\text{Intensity} = A_1 \exp^{-0.5 \left( \frac{\nu - \nu_1}{\Delta \nu_1} \right)^2} + A_2 \exp^{-0.5 \left( \frac{\nu - \nu_2}{\Delta \nu_2} \right)^2} \tag{3}
\]
Images were representative of the real sample characteristics. To image samples described in "Bottle tests." Each type of sample was imaged at least 10 times, to ensure that shifts to higher wavenumbers (blueshifts) would indicate an enhancement in the structure of water.

Optical microscopy. An optical VHX-5000 digital microscope (Keyence Corporation, Canada) was used to image samples described in "Bottle tests." Each type of sample was imaged at least 10 times, to ensure that images were representative of the real sample characteristics.

Flow experiments through injectable filters. Injectable filters were obtained as described elsewhere. Briefly, a glass graduated cylinder having a 14 mm diameter was perforated at the bottom and filled with sand for a total volume of 5 mL, to mimic an aquifer. Water (20 mL) was flushed through the cylinder before injecting 1 mL of 0.1% wt HEC + solutions in water, followed by rinsing with excess water (60 mL). We subsequently injected in the cylinder co-contaminant mixtures of copper salts and toluene (samples 3 and 4 described in Table 1, "Static interfacial tension measurement"). Note that samples filtered through injectable filters were not allowed to settle overnight. The eluent was collected from the bottom of the cylinder and analyzed using ICP-OES (as described in "ICP-OES"), to assess copper concentrations in the water phase. Experiments were conducted in duplicate.

ICP-OES. ICP-OES tests were conducted in triplicate using a PerkinElmer 5300 DV instrument, to quantitate the removal of copper ions from CuCl₂ and CuSO₄ aqueous solutions, using lauric acid. The summary of the samples analyzed is provided in Table 2. Bottle test samples were allowed to settle overnight, before collecting the water phase. The eluent from flow experiments described in “Flow experiments through injectable filters” was not allowed to settle overnight (i.e., samples were filtered immediately after mixing and the eluent was immediately collected). Each measurement was conducted in triplicate.

Synchrotron based Small Angle X-ray Scattering (SAXS). SAXS experiments were carried out at the Canadian Light Source Synchrotron (CLS) on the Brockhouse Diffraction Sector Undulator Beamline (BXDS-IVU), to identify the self-assembly of lauric acid (500 g/L) into crystal structures into pure THF and in THF-water mixtures. Samples were prepared using water at circum-neutral and acidic pH (pH = 2) and with different THF-water ratios (90–95% THF relative to water, v/v). Samples were pipetted and sealed in a Kapton tubing (Polyamide 0.0575" ID × 0.0615" OD) with wax to prevent solvent evaporation during data collection. SAXS data was collected with a Rigaku MX300HE detector (8192 × 8192 pixels) with 2 × 2 binning (4096 × 4096 pixel) for an effective pixel size of 73.242 μm with background stability mode ON. SAXS patterns were collected with a photon energy of 12.18 keV and sample-to-detector distance of ~233 cm. SAXS patterns were collected in transmission geometry with a 10 s dwell time. Patterns were processed with GSASII (Argonne National Laboratory, C, 2010) This product includes software developed by the UChicago Argonne, LLC. SAXS data was calibrated with silver behenate (AgBeh) and instrument parameters such as sample-to-detector distances, detector tilt, beam centre, were refined as described elsewhere. SAXS patterns were integrated from q = 0.012 to q = 0.637 Å⁻¹. Measurements were duplicated for each sample.

The scattering intensity was normalized [0,1] and fitted by nonlinear regression with Graphpad Prism 9.4.0 (Graphpad software, San Diego, California). Prism uses the Levenberg and Marquardt algorithm for non linear regression. We fitted three different models to the data: Gaussian, Teubner and Strey, and Ornstein–Zernike. The Gaussian model is given by Eq. (4).
where \( I(q) \) is the scattering intensity (expressed as a function of the scattering vector \( q \)), \( A \) is the amplitude (\( A = 1 \) in this case, since the data were normalized), \( \sigma \) is the standard deviation, and \( \mu \) is the mean.

The Teubner and Strey model\(^8\), shown in Eq. (5), was fitted to the normalized scattering intensity vs. \( q \) data:

\[
I(q) = \frac{C}{a^2 + c_1 q^2 + c_2 q^4}
\]

where \( C \) is a proportionality constant, set to 1 in our treatment due to data normalization, and \( a, c_1 \) and \( c_2 \) are fitting parameters, which were used to determine the periodicity \( d \) and the correlation length \( \xi \), as shown in Eqs. (6)–(7).

\[
\xi = \left[ \frac{1}{2} \left( \frac{a^2}{c_2} \right)^{1/2} + \frac{1}{4} \frac{c_1}{c_2} \right]^{-1/2}
\]

\[
d = 2\pi \left[ \frac{1}{2} \left( \frac{a^2}{c_2} \right)^{1/2} - \frac{1}{4} \frac{c_1}{c_2} \right]^{-1/2}
\]

The Ornstein–Zernike model was fitted to normalized scattering intensity vs. \( q \) data using Eq. (8).

\[
I(q) = \frac{1}{a^2 + c_1 q^2}
\]

**\(^1\)H NMR.** \(^1\)H NMR was used to analyze the purity of the water-rich phase separated from samples prepared with 0.125 M or 1 M lauric acid and mixtures of 7:3 water:THF (v/v). Samples were transferred directly into 5 mm NMR tubes without the addition of any deuterated solvent. NMR spectra were collected on a Bruker AVANCE NEO 300 MHz spectrometer equipped with a 5 mm BBFO probe. The sample temperature was regulated at 298 ± 1 K. Quantitative spectra were collected with the field frequency lock turned off, using a single scan with a relaxation delay of 30 s and an acquisition time of 5.6 s. Spectra were processed with 0.3 Hz line broadening and polynomial baseline correction.

Under quantitative NMR conditions, the integral of an NMR peak is proportional to the number of protons contributing to the peak (e.g., two for \( H_2O \) and four for THF). Thus, in a two-component THF-water system, the mole fraction of one component (e.g., water in the water-rich phase) is given by Eq. (9)

\[
\chi_{water} = \frac{I_{water}}{I_{water}^2 + I_S^4}
\]

where \( I_{water} \) and \( I_S \) are the NMR peak integrals of water and either THF, respectively. Note that the integration limits for the organic solvents encompass both the main peak and the \(^{13}\)C satellite peaks.

The above equation is only true for a two-component system. While the NMR spectra did reveal the presence of lauric acid in the sample, the mole fraction of lauric acid was less than 0.0003 in the water-rich fraction of all samples. As such, the water-rich phase of all samples is effectively a two-component system for the purpose of this analysis.

The partial molar volumes of THF and water\(^8\) were used to compute %v/v water from \( \chi_{water} \). Each sample was analyzed in triplicate.

**Light scattering.** Light scattering experiments were conducted in triplicate at room temperature (23 °C) using a Malvern Zetasizer Nano ZSP. Samples were equilibrated for 120 s prior to each measurement and measurements were repeated at least three times for every sample. The refractive index used was 1.3 for water\(^18\),\(^87\) and 1.423 for lauric acid. The density, viscosity and refractive indexes of THF-water mixtures depended on the THF:water ratio, and are as reported elsewhere\(^88\),\(^89\). Measurements were conducted in triplicate.

**Results and discussion**

This section discusses emulsification between toluene and water, and separation between water and THF in the presence of lauric acid (“Emulsification and solvent separation by lauric acid (without Cu\(^{2+}\)”). We also discuss the mechanisms through which copper ions migrate in the organic phase following solvent separation, in the presence of lauric acid (“Partitioning of Cu\(^{2+}\) ions between the organic solvent and the water phase”). We analyze how copper ions decrease the stability of water-toluene and THF-water emulsions, promoting the separation between free water-rich and organic solvent-rich phases, where copper ions partition (“Emulsion stability with copper salts”). Finally, “Flow test experiments (injectable filters)” discusses injectable filters, which can be used in conjunction with lauric acid for the treatment of copper and toluene pollution in groundwater.
Emulsification and solvent separation by lauric acid (without Cu²⁺). Lauric acid emulsifies toluene and water, and emulsion stability increases from pH = 2 to pH = 6.5, and from pH = 6.5 to pH = 13 (Fig. S1, supporting information file). Emulsion stability at basic pH is correlated to low interfacial tension. The interfacial tension measured at the toluene-water interface, using 0.125 M lauric acid (relative to the toluene phase) is similar at acidic pH (20.0 ± 1.1 mN/m) and at circum-neutral pH (24.2 ± 2.7 mN/m), and lower at basic pH (13.0 ± 0.4 mN/m).

The pKa of lauric acid is ≈5.90. A study conducted in pure water reports that at pH < 4, lauric acid is not dissociated, at circum-neutral pH a portion of lauric acid is in the form of laurate ions, and at pH > 9 all of lauric acid in aqueous solution is in the form of laurate ions. While lauric acid is insoluble in water, laurate ions have greater affinity for the water phase. Interfacial tension measurements show that laurate ions are more interfacially active and more effective than lauric acid at stabilizing toluene-water emulsions.

Other studies conducted using lauric acid, water and water-immiscible oils also report that the pH of the water phase affects emulsification. For instance, a study conducted using solutions of lauric acid in paraffin and mineral oil reports that sodium laurate films form at the oil–water interface with NaOH > 0.2 M. These films stabilize droplets because of their negative electrostatic charge at alkaline pH. Another study reports that mixed films of lauric acid and laurate enhance the stability of water-cyclohexane and water-n-hexadecane emulsions at basic pH. Moreover, the oil:water ratio affects emulsion characteristics. As an example, a previous study reports that at high water concentrations, micelles incorporate oil in the aqueous surfactant solution, whereas at high oil:water ratios micelles incorporate water in oil. Here, the toluene:water ratio used (7:3 toluene:water, v/v) enables the formation of both oil in water and water in oil emulsions at basic pH, as evident from the turbidity of both the toluene and the water layers.

In addition to emulsifying toluene and water, lauric acid also separates THF from water (Fig. 1 and Fig. S2, supporting information file). The purity of the THF and water phases is given in Table 3 (as determined by 1H NMR). The data show that the pH has negligible effect on the purity of the separated water phase after 24 h equilibration. In contrast, lauric acid concentration affect water purity. At circum-neutral pH, purity of the water-rich phase increases with increasing lauric acid concentration (from 0.125 to 1 M, Fig. 2). Nonetheless, using excessively high lauric acid concentrations would not be feasible for practical purposes, and we therefore focus on lower concentrations (e.g., 0.125 M).

We hypothesize that THF-water separation starts with water partitioning inside self-assembled lauric acid structures. SAXS measurements were conducted to characterize these structures, using THF and THF-water mixtures (containing 90% and 95% THF relative to water, v/v, at circum-neutral and acidic pH, Fig. 3). SAXS patterns are characterized by a single peak, which was fit using the model developed by Teubner and Strey for reverse micelles, which are a thermodynamically stable fluid comprised of water, a hydrophobic solvent, and an amphiphile. This is different from a kinetically stable emulsion, which separates into a hydrophobic solvent phase and a water phase over time. Samples used in SAXS measurements had a water- to-surfactant molar ratio of 2.2 or less.
In the Teubner and Strey model, the intensity is described by Eq. (2), as reported in the "Materials and methods" section. For the purpose of the fit, in our study the sign in the original expression by Teubner and Strey was substituted with an equal sign in Eq. (2). Note that in Eq. (2) we set the proportionality constant $C = 1$, because normalized data were used.

The Teubner and Strey model is derived from the phenomenological model developed by Landau to describe changes in free energy $F(\eta)$ during phase transitions (where $\eta$ is the order parameter). When a system is not undergoing transitions, $\eta$ is constant. The free energy $F(\eta)$ is obtained as the integral of the free energy density $f(\eta)$. Near the phase transition, changes in $\eta$ are small and $f(\eta)$ can therefore be approximated by a truncated Taylor expansion in $\eta$. The coefficients of the Taylor expansion are the fitting parameters which appear in Eq. (2). The order at which the expansion is truncated depends on the problem analyzed and the accuracy desired. Note the difference between the Teubner and Strey model (which accounts for terms up to $q^4$) and the Ornstein–Zernike model (Eq. 8), which is obtained when considering only the components corresponding to small order parameter fluctuations and only accounts for terms up to $q^2$.

The Ornstein–Zernike equation was developed by Leonard Ornstein and Frits Zernike. The Ornstein–Zernike equation has been used to estimate the structure factor of liquids and colloids in scattering experiments. While it has been used to study liquids, it is not suitable to describe reverse micelles, as reported by Teubner and Strey and as shown in Fig. 3.

The periodicity $d$ and the correlation length $\xi$ for the Teubner and Strey model are provided in Table 4, alongside with the fitting parameters for the samples analyzed.

Based on the Teubner and Strey model, lauric acid self-assembles into micelles, the size of which increases with increasing water content, at neutral pH, as indicated by the increasing value of $\xi$. This result is attributed to the fact that as the water content increases, the amount of water partitioned into the reverse micelles also increases. With 90% THF and 10% water, the reverse micelle size is greater at neutral than at acidic pH, at which the hydrophilic heads of lauric acid would have a lower negative charge density, leading to decreased electrostatic

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**Table 3.** $^1$H NMR analysis of the water-rich separated layer, obtained starting from 30% v/v THF-70% v/v DI water mixtures, and either 0.125 M or 1 M lauric acid. Samples were centrifuged before analyzing each layer.

| Sample pH       | THF in water layer (% v/v) | Water in water layer (% v/v) |
|-----------------|---------------------------|-----------------------------|
| Acidic, 0.125 M | 24.2 ± 0.4                | 75.8 ± 0.4                  |
| Acidic, 1 M     | 13.4 ± 0.6                | 86.6 ± 0.6                  |
| Circum-neutral, 0.125 M | 24.2 ± 0.4            | 75.8 ± 0.4                  |
| Circum-neutral, 1 M | 13.6 ± 0.2             | 86.4 ± 0.2                  |
| Basic, 0.125 M  | 24.5 ± 0.5                | 75.5 ± 0.2                  |
| Basic, 1 M      | 14.0 ± 0.1                | 86.0 ± 0.1                  |

**Figure 2.** Separation between THF and water at circum-neutral pH (p$\text{H} = 6.5$), with different lauric acid concentrations ranging from 0.2 to 1 M (as indicated on the x axis). Volumes were measured after centrifuging samples for 10 min at 3500 RPM. The THF percentages reported in the legend are relative to the water phase (v/v). The deviation on the ordinate axis was measured using Eq. (1) as follows ($V_{\text{water,measured}} - V_{\text{water,used}} / V_{\text{water,used}}$)*100 (%).
Figure 3. SAXS patterns of lauric acid in either 100% THF, or in THF-water mixtures, with 95% THF or 90% THF (at either acidic or circum-neutral pH) relative to water (v/v) and 500 g/L lauric acid. The experimental data are shown in red, and fit to the Steubner and Strey model (continuous black line) or a Gaussian (dashed line). The Ornstein and Zernicke fit (dotted line) is poor, and shown as a reference. The best fitting parameters of the Ornstein and Zernicke model are as follows: 100% THF: $a_2 = 0.6572$, $c_1 = 152.8$; 95% THF, neutral pH: $a_2 = 0.6366$, $c_1 = 167.0$; 90% THF, neutral pH: $a_2 = 0.6402$, $c_1 = 173.0$; 90% THF, acidic pH: $a_2 = 0.6249$, $c_1 = 175.9$. The Teubner and Strey best fitting parameters are as indicated in Table 4.

### Table 4

| Fitting parameters | $\xi$ (Å) | $d$ (Å) | $d/\xi$ |
|-------------------|------------|---------|---------|
| 100% THF | $a_2$ | 1.401 | 24.3 | 84.8 | 3.48 |
| | $c_1$ | $-2.069 \times 10^2$ | | |
| | $c_2$ | $27.258 \times 10^3$ | | |
| 95% THF, neutral pH | $a_2$ | 1.419 | 25.4 | 87.4 | 3.44 |
| | $c_1$ | $-2.272 \times 10^2$ | | |
| | $c_2$ | $31.412 \times 10^3$ | | |
| 90% THF, neutral pH | $a_2$ | 1.461 | 26.4 | 89.0 | 3.36 |
| | $c_1$ | $-2.522 \times 10^2$ | | |
| | $c_2$ | $35.469 \times 10^3$ | | |
| 90% THF, acidic pH | $a_2$ | 1.409 | 25.8 | 89.7 | 3.48 |
| | $c_1$ | $-2.335 \times 10^2$ | | |
| | $c_2$ | $34.310 \times 10^3$ | | |

Table 4. Periodicity $d$ and to the correlation length $\xi$ estimated using the Teubner and Strey model for different samples containing lauric acid (500 g/L) in either THF (100% THF) or THF-water mixtures (with 90% or 95% THF relative to water, v/v), at neutral and acidic pH. The fitting parameters for the samples are also provided. Note the negative values of $c_1$, typical of reverse micelles.85
repulsion and tighter packing. Previous atomistic simulation studies report that the self-assembly of lauric acid in pure water is pH-dependent\textsuperscript{91,97}. At pH = 13 it self-assembles into micelles, whereas at pH = 4–9 it self-assembles into bilayers, and the order parameter increases with decreasing pH\textsuperscript{91,97}. In our study, the solvent was either THF or a mixture of THF and water. Organic solvents in solution would induce the self-assembly of lauric acid into reverse micelles, as reported in a study which used it to remove methylene blue from water, by causing its migration into isopentanol (used as the organic solvent)\textsuperscript{98}.

The periodicity $d$ determined based on the Teubner and Strey model is $\approx 3.4$ times the correlation length $\xi$, indicating that reverse micelles of lauric acid aggregate. The periodicity $d$ is interpreted as the average distance between the centers of lauric acid reverse micelles within the aggregates. Aggregation between reverse micelles is ascribed to the high lauric acid concentrations used in SAXS measurements.

Note that while a Gaussian model provides a good mathematical fit to the data (Fig. 3), it does not adequately capture the physical characteristics of microemulsions. Based on a Gaussian model, the micelle size is estimated as $2\pi/q_{\text{max}}$ (where $q_{\text{max}}$ is the position of the peak maximum) and is 109–115 Å for the different samples analyzed (Table S4, supporting information file). This size is larger than the one determined with Teubner and Strey, and does not correlate well to the size of lauric acid molecules.

Light scattering experiments were conducted to complement SAXS results, to probe different length scales. With 90% DMSO, droplet size was above the range of detection of the zeta-sizer used for light scattering experiments, and samples phase separated after 48 h at either circum-neutral or acidic pH. With 95% THF, light scattering experiments reveal that droplets are $450 \pm 82$ nm and $505 \pm 176$ nm at circum-neutral and acidic pH, respectively.

In summary, based on our results we hypothesize that reverse micelles are the precursors to droplets, and that they grow upon increasing the water content in THF, in agreement with previous studies\textsuperscript{99}. Reverse micelles would grow into droplets, and progressive increase in droplet size would ultimately lead to phase separation (Fig. 4). Recall that bottle tests show that at circum-neutral pH, emulsions are water in THF (rather than THF in water). Therefore, the data suggest that when the percentage of water exceeds the percentage of THF, reverse micelles swell to host increasing volumes of water in their interior and partition in the THF-rich phase.

ATR-FTIR was used to analyze intermolecular interactions between THF, water and lauric acid. This analysis was conducted to probe if lauric acid interfered with THF-water interactions, inducing solvent separation. THF and water interact through H-bonds\textsuperscript{100}. Species such as sugars and sodium lauroyl lactylate\textsuperscript{64} preferentially interact with water through H-bonds, causing separation between water and THF or acetonitrile\textsuperscript{76}. At circum-neutral pH, part of lauric acid should be in the form of laurate ions (as discussed earlier). Although laurate ions are interfacially active, analysis of the H-bond region indicates that lauric acid had a negligible effect on H-bonding of the water phase at either acidic or circum-neutral pH (Fig. S3 and Tables S1–S3, supporting information file). Since water-THF separation by lauric acid cannot be explained based on the effect of lauric acid on H-bonding of water, further ATR-FTIR analyses were conducted on samples containing 1 M lauric acid in 3:7 THF:water (v/v), at circum-neutral, acidic and basic pH. Specifically, we analyzed the top and bottom phase of the samples, to determine the purity of the separated phases and to probe the effect of pH on lauric acid partitioning between phases. The spectra of all THF-water mixtures (without lauric acid) were largely similar at all pH values analyzed (Fig. S5, supporting information file). The spectra of the bottom water-rich phase were also spectroscopically identical at all pH values (Fig. S6, supporting information file). Moreover, the spectra of the bottom layers of THF-water mixtures separated by lauric acid and of THF-water mixtures without lauric acid are largely similar, with the exception of the intensity of the THF band in the 2900 cm$^{-1}$ region (inset of Fig. S6, supporting information file). This result indicates that the water-rich layer separated by lauric acid does not contain lauric acid (i.e., lauric acid partitions mainly in the THF-rich phase). The intensity difference seen in the CH stretching region can be used to roughly estimate the THF content in the bottom water-rich layer. Integrating this peak and comparing

![Figure 4](https://doi.org/10.1038/s41598-022-20241-4)
its area to the one of the peak measured in 3:7 THF:water mixture (before separation with lauric acid), we determined that the THF content in the bottom layer is 18.5%, in agreement with 1H NMR data.

Figure 5 shows the spectra of 3:7 THF:water mixtures (without lauric acid) and of the top layer of mixtures separated by lauric acid, at different pH values. Tables 4 and 5 highlight relevant absorbance peaks. The spectrum of pure water is also included for comparison. The spectra of lauric acid powder, pure THF and 1.5 M lauric acid in THF are shown in Fig. S4 (supporting information file), as a reference. The broad band at wavenumbers greater than 3000 cm⁻¹ corresponds to O–H stretching vibrations. The peaks between 3000 and 2800 cm⁻¹ are due to the CH stretching vibrations of both THF and lauric acid. The double peak centred at 1720 cm⁻¹ corresponds to the C = O stretching mode of lauric acid, specifically to lauric acid molecules with two hydrogen bonds (low frequency) and with one hydrogen bond (high frequency). The water bending mode (δHOH) is at 1640 cm⁻¹.

At lower frequencies, the fingerprint region contains a number of peaks corresponding to characteristic vibrations, including aliphatic bending (at 1460 cm⁻¹) and C–O stretching (at ~ 1060 cm⁻¹) modes. Note that the spectra do not show the characteristic peaks of carboxylates around 1580–1550 cm⁻¹ (for the antisymmetric COO⁻ vibrations) and 1417–1308 cm⁻¹ (for the symmetric COO⁻ vibrations). This indicates that the top layer emulsion contains lauric acid in its protonated form only. Although the spectra of the bottom water-rich layer are identical for all pH values, the intensity of the top-layer carbonyl stretching band at 1720 cm⁻¹ decreases with decreasing pH. Therefore, we hypothesize that the deprotonated form of lauric acid preferentially partitions at the interface between the water-rich and the THF-rich layers. This hypothesis is further supported by our interfacial tension measurements, as will be discussed later in this section.

Wu et al. used the shift of the water bending band 1640 cm⁻¹ (inset, Fig. 5) to quantify the water content in mixtures of water and glycerol, with higher frequencies corresponding to lower water content. Table 5 shows the δHOH frequency for each sample. These values indicate that the water content increases with pH and in all cases is lower than in the 3:7 THF:water (v/v) mixture (without lauric acid).
This observation is further supported by analysis of the O–H stretching region. Indeed, the area under the O–H stretching band increases with increasing pH, indicating that more water is present in the THF phase at higher pH. However, the shape of the band is identical for all pH values, indicating that the nature of H-bonds is the same at all pH values. Hence, differences between separation efficiency and emulsion stability are not due to changes in H-bonding in the bulk THF and water phases.

The C–O stretch peak of THF for each of the samples also undergoes a shift, as indicated by the grey dashed line in Fig. 5. The higher frequency shoulder centred at 1068 cm⁻¹ corresponds to the C–O stretch from the carboxylic group of lauric acid, whereas the lower frequency peak ranging from 1035 to 1040 cm⁻¹ corresponds to the C–O stretch of the THF. Purkayastha and Madhurima show the spectra of progressively hydrated THF, where the C–O stretch band shifts to lower frequencies as a function of hydration. In our case, the C–O stretch band shifts from 1066 cm⁻¹ for pure THF to 1040 cm⁻¹ for 3:7 THF:water mixtures without lauric acid. The shift seen in the absorbance peaks of ATR-FTIR spectra is summarized in Table 6. The data indicate that that the top layer of separated THF-water mixtures is enriched in THF compared to 3:7 THF:water mixtures without lauric acid. The shift in the spectra is associated with changes in the nature of the hydrogen bonds within the THF, as indicated by the change in the frequency of the C–O stretch band.

Partitioning of Cu²⁺ ions between the organic solvent and the water phase. As discussed in the previous section, lauric acid separates THF and water. Separation between THF and water also occurs with aqueous solutions of either 0.03 M CuSO₄ or 0.03 M CuCl₂ (Fig. 6). With 30% THF (relative to the water phase, v/v), 0.125 M lauric acid and at circum-neutral pH, the percent difference (deviation) between the initial water added and the volume of the water-rich layer after separation is 28.1 ± 1.0% and 28.1 ± 1.0% with 0.03 M CuCl₂ and 0.03 M CuSO₄, respectively. Increasing lauric acid concentration to 0.25 M decreases the deviation to approximately 25% with aqueous solutions of either 0.03 M CuSO₄ or 0.03 M CuCl₂. This result is similar to observations without copper salts (Fig. 2), which show increased THF-water separation with increasing lauric acid concentrations.

Upon THF-water separation, Cu²⁺ ions partition in the THF phase at circum-neutral pH, as qualitatively observed based on the discoloration of the top THF-rich layer (Fig. 6). This does not occur at pH = 2 (at which

| Sample                                      | Frequency C–O stretch/cm⁻¹ |
|---------------------------------------------|-----------------------------|
| Pure THF                                    | 1066                        |
| Top layer—acidic                            | 1056                        |
| Top layer—circum-neutral                    | 1053                        |
| Top layer—basic                             | 1050                        |
| 3:7 THF:water mixture (without lauric acid) | 1041                        |

Table 6. Frequency of the C–O stretch of THF for each sample, prepared either with or without 1 M lauric acid. In the table, the rows labelled as 'top layer' refer to the top layer of 3:7 THF:water (v/v) mixtures separated with 1 M lauric acid.
Cu²⁺ remains dissolved in water, Fig. S9, supporting information file). At pH = 13, part of the copper ions partition in the THF layer, and part precipitate out of solution, entraining laurate ions (as evident from the white color of the precipitates, Fig. S10, supporting information file). Note that precipitation of Cu²⁺ ions at basic pH would also occur without lauric acid. These results show that circum-neutral pH is preferable to achieve the simultaneous separation of THF and copper ions from water, when using lauric acid.

The data show that laurate ions (present at circum-neutral pH) act as carriers for Cu²⁺, transporting it into the THF phase. Cu²⁺ remain in the THF-rich phase over a period of approximately three weeks, after which they return into the water phase. These results indicate that Cu²⁺ migration into the THF-rich phase is a transient phenomenon, which would however allow removal of copper from water. We conducted ATR-FTIR experiments to probe copper-lauric acid interactions. Figure 7 shows the ATR-FTIR spectra of the top and bottom phase of samples prepared with 1 M lauric acid and either 3:7 THF:100 mM CuSO₄ mixtures (using water at circum-neutral pH) or 3:7 THF:water mixtures (using water at acidic pH). The reason for using water at acidic pH in the absence of copper salts is to have a largely clear top, THF-rich layer. The top layer is also largely clear at circum-neutral pH, in the presence of copper (Fig. 6).

**Figure 6.** Bottle tests conducted using aqueous solutions of 0.03 M CuCl₂ and CuSO₄, using 30% or 50% THF v/v relative to water, at circum-neutral and acidic pH. Copper ions partition in the THF phase only at circum-neutral, while they remain solubilized in water at pH = 2 and precipitate in part out of solution at pH = 13 (Fig. S10, supporting information file). Images were taken after leaving vials to sit overnight.

**Figure 7.** ATR-FTIR spectra of samples prepared with 1 M lauric acid, and 3:7 THF: 100 mM CuSO₄ in water, at circum-neutral and acidic pH. These samples were separated into a top and a bottom phase. The color coding is as follows: bottom layer (black), top layer (red) and top layer at acidic pH (green). Red and green lines are identical in shape and intensity. Inset shows the sulfate absorbing region, indicating that there is no sulfate in the top layer of separated THF-water mixtures, in the presence of copper.
Copper partitioning between toluene and water was examined through bottle tests conducted using 30 mM CuCl₂ or CuSO₄ solutions, toluene, and either lauric acid alone or a solution of lauric acid in canola oil (added as a 23 mM lauric acid in canola oil solution), when using water with initial pH = 13 (adjusted with 10 mM NaOH). Dispensing lauric acid in canola oil solutions would be useful for groundwater remediation, to deliver lauric acid or Cu₂⁺ solutions in toluene and lauric acid solutions in canola oil. The volumes used and the lauric acid concentrations are as indicated in the table. Partitioning of copper ions in the toluene phase with lauric acid and a base is also achieved without using canola oil, as qualitatively assessed based on the discoloration of the toluene phase (Fig. S13, supporting information). Measurements were done after 24 h equilibration.

**Table 7.** ICP analyses of the water phase of samples prepared starting with 30 mM CuCl₂ and 10 mM CuSO₄, using toluene and lauric acid solutions in canola oil. The volumes used and the lauric acid concentrations are as indicated in the table. Partitioning of copper ions in the toluene phase with lauric acid and a base is also achieved without using canola oil, as qualitatively assessed based on the discoloration of the toluene phase (Fig. S13, supporting information). Measurements were done after 24 h equilibration.

| Water phase (5 mL) | Oil phase (total volume in brackets) | Cu²⁺ concentration in water after treatment (mM) [Average % Cu²⁺ removal] |
|--------------------|--------------------------------------|------------------------------------------------------------------------|
| 30 mM CuCl₂ + 10 mM NaOH | 1:1 toluene: 0.25 M lauric acid in canola oil solution (2 mL) | 17.2 ± 2.0 [42.8] |
| 10 mM CuCl₂ + 10 mM NaOH | 1:1 toluene: 0.25 M lauric acid in canola oil solution (2 mL) | 3.9 ± 0.1 [61.1] |
| 30 mM CuCl₂ + 10 mM NaOH | 1:4 toluene: 0.25 M lauric acid in canola oil solution (5 mL) | 13.3 ± 1.1 [55.7] |

**Table 8.** ICP analyses of the water phase of samples prepared starting with 30 mM CuSO₄ and 10 mM CuSO₄, using toluene and lauric acid solutions in canola oil. The volumes used and the lauric acid concentrations are as indicated in the table. Partitioning of copper ions in the toluene phase with lauric acid and a base is also achieved without using canola oil, as qualitatively assessed based on the discoloration of the toluene phase (Fig. S13, supporting information). Measurements were done after 24 h equilibration.

| Water phase (5 mL) | Oil phase (total volume in brackets) | Cu²⁺ concentration in water after treatment (mM) [Average % Cu²⁺ removal] |
|--------------------|--------------------------------------|------------------------------------------------------------------------|
| 30 mM CuSO₄ + 10 mM NaOH | 1:1 toluene: 0.25 M lauric acid in canola oil solution (2 mL) | 17.6 ± 2.5 [41.4] |
| 10 mM CuSO₄ + 10 mM NaOH | 1:1 toluene: 0.25 M lauric acid in canola oil solution (2 mL) | 4.2 ± 0.2 [57.9] |
| 30 mM CuSO₄ + 10 mM NaOH | 1:4 toluene: 0.25 M lauric acid in canola oil solution (5 mL) | 15.4 ± 2.0 [48.8] |

Copper (inset shows sulfate region in Fig. 7). Also, the signatures of Cu-laurate complexes and sulfates are not detected in ATR-FTIR spectra. Nonetheless, the top layer has a blue discoloration, as highlighted above (Fig. 6). Therefore, the data suggest that laurate ions release free copper ions in the top THF-rich phase, rather than forming stable complexes with it. Previous studies report that metal ion transport into organic phases can be mediated by carboxylate ions\(^{[15-16]}\), in agreement with our hypothesis and results. After dissociating from laurate ions, Cu²⁺ ions eventually return to the water phase, as indicated earlier.

Finally, before discussing Cu²⁺ partitioning into water-immiscible solvents such as toluene and hexane, note that Cu²⁺ also partitions into dioxane at circum-neutral pH (although lauric acid-induced dioxane-water separation is poor, Figs. S8–S11).

Copper partitioning between toluene and water was examined through bottle tests conducted using 30 mM CuCl₂ or CuSO₄ solutions, toluene, and either lauric acid alone or a solution of lauric acid in canola oil (added as a 23 mM lauric acid in canola oil solution), when using water with initial pH = 13 (adjusted with 10 mM NaOH). Dispensing lauric acid in canola oil solutions would be useful for groundwater remediation, to deliver lauric acid to the polluted zones. When treating water at the surface, neat lauric acid can be used instead. Bottle tests show that Cu²⁺ initially partitions in the toluene phase in the presence of lauric acid, when using water with initial pH = 13 (adjusted with NaOH, Figs. S12, S13, supporting information file). After 24 h, Cu²⁺ ions eventually return to the water phase, as indicated earlier. But basic pH, lauric acid was dissociated and it would therefore be able to interact with copper cations (e.g., through electrostatic interactions).

ICP tests conducted on the water phase was used to quantify the percent removal of Cu²⁺ ions from CuCl₂ and CuSO₄ solutions (Tables 7, 8). Removal was greatest (~60%) with initial concentrations of 10 mM copper ions in water, when using 1:1 0.25 M lauric acid in canola oil:toluene as the oil phase. Percent removal efficiency decreases with 10 mM copper ions in water, when the lauric acid concentration is kept constant. This is likely because less copper can bind to lauric acid, when the ratio of lauric acid to copper ions decreases. Future research will focus on identifying compounds which enable greater copper ion removal from water.

As discussed earlier, Cu²⁺ and laurate ions likely interact through physical interactions (e.g., electrostatic interactions), causing Cu²⁺ to partition at the toluene-water interface. The ATR-FTIR spectra of the oil phase of samples containing either 30 mM CuCl₂ or CuSO₄ and 10 mM NaOH (relative to the water phase), lauric acid in canola oil and toluene are similar to those of samples containing lauric acid in toluene, with the exception of the peak at ~1160 cm⁻¹ (Fig. 8). This peak is also observed in the top phase of samples prepared with water and 10 mM NaOH, 0.25 M lauric acid in canola oil and toluene, without Cu²⁺. Therefore, it is not representative of chemical bonding between Cu²⁺ and lauric acid or laurate ions. New peaks do not appear in the spectra of
samples containing both lauric acid and copper salts. Therefore, ATR-FTIR data do not highlight any chemical bonding between Cu²⁺ and lauric acid, in agreement with the results obtained with THF.

Note the marked ATR-FTIR -OH scissor and -OH stretch absorbance peaks at ≈1600 cm⁻¹ and ≈3400 cm⁻¹ in the ATR FTIR spectra of samples prepared with water and 10 mM NaOH, 0.25 M lauric acid in canola oil and toluene (without copper salts). These peaks are absent in the oil phase of samples which also contain copper salts, indicating negligible water content. This result indicates that water in toluene emulsions are more stable with than without copper salts. Emulsion stability is further discussed in the following section (“Emulsion stability with copper salts”).

Emulsion stability with copper salts. The goal of this study is to use lauric acid to separate water from toluene and copper, and THF and copper. Stable emulsions of either THF or toluene in water would entrain organics solvents and copper (partitioned inside them) into the water phase. Unless filters are used, this would hinder water purification. “Flow test experiments (injectable filters)” will discuss injectable filters, which can be used in conjunction with lauric acid for the treatment of copper and toluene pollution in groundwater. Stable water in oil emulsions would be less problematic, although they would lead to larger volumes of liquid waste following treatment.

Water in THF and THF in water emulsions are not stable with THF and either CuCl₂ or CuSO₄, as shown for instance in Fig. 6 (where mixtures are separated into free phases). As a result, there should be limited entrainment of Cu²⁺ in the water phase.

With 50% water and 50% toluene (without canola oil), some toluene remains emulsified in water after a day, as indicated by residual turbidity of the water phase (Fig. S13, supporting information file). Similar results are obtained with canola oil (Fig. S12, supporting information file). Recall that canola oil is used to enable the delivery of lauric acid into polluted aquifers. Over short time intervals, water in oil and oil in water emulsions are observed for samples prepared with either 30 mM CuCl₂ or CuSO₄ aqueous solutions, toluene and lauric acid in canola oil solutions (cf. optical microscopy images, Fig. 9). Toluene droplets in water can be excluded using filters, as we will further discuss in the following section (“Flow test experiments (injectable filters)”). Water in oil emulsions are promptly destabilized with CuSO₄, whereas sparse oil in water droplets are observed with CuCl₂ (image not shown). After 24 h, water in oil emulsions are also not observed with CuCl₂. After this time period, flocs segregate at the oil–water interface (e.g., Fig. S14, supporting information).

The interfacial tension measured with lauric acid at the water–canola oil + toluene interface is 10.30 ± 0.18 mN/m, with 10 mM NaOH. The interfacial tension of lauric acid decreases to 27.32 ± 0.03 mN/m with CuSO₄ and NaOH, confirming that Cu²⁺ interacts with lauric acid (likely through electrostatic interactions), forming flocs which are less effective at stabilizing emulsions than free laurate ions.

Compressional rigidity is also correlated to emulsion stability. Compressional isotherms measured with lauric acid and CuSO₄ are largely flat (data not shown), indicating that either films are soft or that interfacially active species rapidly desorb upon compression. Either scenario would lead to poor emulsion stability. Note that other factors also affect emulsion stability. As an example, flocs can contribute steric repulsion between...
droplets and stabilize them through Pickering stabilization mechanisms. Pickering stabilization mechanisms can account for the stability of emulsions observed with lauric acid and CuSO₄ over short time periods. Compression isotherms measured with lauric acid and CuCl₂ indicate that interfacial films are rigid, as demonstrated by the increase in pressure (19 ± 3 mN/m, Fig. 10). When films were re-compressed immediately after the first compression, the re-compression isotherm is similar to the first compression. This indicates that interfacially active species either do not desorb upon compression, or they rapidly re-adsorb. Either scenario would promote emulsion stability, explaining why emulsions were more stable with CuCl₂ than with CuSO₄. Nonetheless, emulsions were not observed after 24 h even with CuCl₂, as highlighted above. Rapid emulsion separation can be achieved even in aquifers using injectable filters, as discussed in “Flow test experiments (injectable filters)”.

Flow test experiments (injectable filters). Flow tests experiments are used to investigate the effectiveness of injectable filters in removing copper from water co-polluted by copper and toluene, when used in conjunction with lauric acid. These filters would enable the simultaneous separation of copper and toluene from water, preventing them from migrating downstream. Pollutants could then be extracted upstream from the filter (Fig. 11).
Our previous studies showed that injectable filters could retain hydrophobic solvents such as toluene and hexane51, and that HEC + injectable filters retained diesel droplets emulsified by bacterial biosurfactants81. Here, HEC + injectable filters retain toluene containing Cu²⁺ ions, as shown by ICP-OES analyses (Table 9). Copper ion removal achieved using filters is comparable to bottle test experiments (≈60%, starting from 30 mM Cu²⁺), even without equilibrating for 24 h. This result indicates that filters aid in the simultaneous removal of copper ions and toluene.

In contrast, these filters cannot exclude THF separated from water by lauric acid, because of its higher polarity compared to toluene. Therefore, lauric acid cannot be used for the in situ separation of water from THF and copper ions. In contrast, lauric acid would be better suited to treat water polluted by copper and THF at the surface, for instance after its extraction from polluted aquifers using pump and treat.

Conclusions
Lauric acid can be used to purify water polluted with THF at ambient temperature (20 °C), with minimal energy costs. At the onset of THF-water separation and with > 90% THF (v/v, relative to water), lauric acid self-assembles into reverse micelles. These reverse micelles are ≈25 Å in size and host water in their interior, as shown by SAXS. They swell with increasing water content, ultimately leading to free phase separation. With 1 M lauric acid and 7:3 THF:water mixtures, the purity of the water phase is 87% at either acidic, semi-neutral or basic pH (as shown by 1H NMR). Separation efficiency decreases at lower lauric acid concentrations, and is 76% with 0.125 M lauric acid. Therefore, our facile treatment approach finds potential applications for the treatment of either wastewater or groundwater.

Lauric acid also removes copper from water co-polluted with either THF or toluene. It interacts with copper ions through physical interactions (e.g., electrostatic interactions) and causes their migration into the organic phase. Model sandy aquifers co-polluted with toluene and Cu²⁺ were treated with lauric acid, and HEC + semi-permeable barriers were placed downstream of the pollutant plume. These barriers retain both toluene and ≈50% Cu²⁺. Therefore, they have the potential to offset the risk of pollutant migration and aid in the protection of downstream receptors. Extraction of co-contaminants upstream of the filters can be achieved using pumping wells, enabling their simultaneous removal from aquifers.

Future research will focus on removing additional ions from water. We will also focus on improving the purity of the water phase separated from water miscible solvents and ions using other amphiphilic molecules.
Data availability

The datasets used and/or analysed during the current study available from the corresponding author on reasonable request.

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References

1. Huang, W.-H., Dong, C.-D., Chen, C.-W., Surampalli, R. Y. & Kao, C.-M. Application of sulfate reduction mechanisms for the simultaneous bioremediation of toluene and copper contaminated groundwater. *Int. Biodeterior. Biodegradation* **124**, 215–222 (2017).
2. Mulligan, C. N., Yong, R. N. & Gibbs, B. F. Remediation technologies for metal-contaminated soils and groundwater: An evaluation. *Eng. Geol.* **60**, 193–207 (2001).
3. Li, J., Alves de Toledo, R. & Shim, H. Multivariate optimization for the simultaneous bioremoval of BTEX and chlorinated aliphatic hydrocarbons by Pseudomonas plecoglossicida. *J. Hazard. Mater.* **321**, 238–246 (2017).
4. Sales, C. M., Grostern, A., Parales, J. V., Parales, R. E. & Alvarez-Cohen, L. Oxidation of the cyclic ethers 1, 4-dioxane and tetrahydrofuran by a monooxygenase in two Pseudonococcia species. *Appl. Environ. Microbiol.* **79**, 7702–7708 (2021).
5. Yin, Y. et al. Progressive freezing and suspension crystallization methods for tetrahydrofuran recovery from Grignard reagent wastewater. *J. Clean. Prod.* **144**, 180–186 (2017).
6. Arijoon, A., Olarain, A. O. & Pillay, R. Co-contamination of water with chlorinated hydrocarbons and heavy metals: challenges and current bioremediation strategies. *Int. J. Environ. Sci. Technol.* **10**, 395–412 (2013).
7. Li, Q. et al. Effects of co-contamination of heavy metals and total petroleum hydrocarbons on soil bacterial community and function network reconstitution. *Ecotoxicol. Environ. Saf.* **204**, 111083 (2020).
8. Chen, F. et al. Interaction of cadmium and polycyclic aromatic hydrocarbons in co-contaminated soil. *Water Air Soil Pollut.* **229**, 1–10 (2018).
9. Sun, Y., Zhou, Q., Xu, Y., Wang, L. & Liang, X. Phytoremediation for co-contaminated soils of benzo[a]pyrene (BaP) and heavy metals using ornamental plant Tagetes patula. *J. Hazard. Mater.* **186**, 2075–2082 (2011).
10. Li, F., Deng, D., Zeng, L., Abrams, S. & Li, M. Sequential anaerobic and aerobic bioaugmentation for commingled groundwater contamination of trichloroethene and 1, 4-dioxane. *Sci. Total Environ.* **774**, 141511 (2021).
11. Syramidou, E., Christofopoulos, S., & Kalogerakis, N. Juncus spp.—the helophyte for all (phyto) remediation purposes? *New Biotechnol.* **38**, 43–55 (2021).
12. Telepanich, A., Marshall, T., Gregori, S., Marangoni, A. G. & Pensini, E. Graphene-Alginic Fluids as Unconventional Electrodes for the Electrokinetic Remediation of Cr(VI) 232 (Air, & Soil Pollution, Water, 2021).
13. Alshawabkeh, A. N. Electrokinetic soil remediation: Challenges and opportunities. *Sep. Sci. Technol.* **44**, 2171–2187 (2009).
14. Gidudu, B. & Chirwa, E. M. N. The combined application of a high voltage, low electrode spacing, and biosurfactants enhances the bio-electrokinetic remediation of petroleum contaminated soil. *J. Clean. Prod.* **276**, 123745 (2022).
15. Kim, S. O., Jeong, J. Y., Lee, W. C., Yun, S. T. & Ho, Y. Y. Electrokinetic remediation of heavy metal-contaminated soils: Performance comparison between one-and two-dimensional electrode configurations. *J. Soils Sedim.* **1**, 1–15 (2020).
16. Kim, W. S. et al. Field application of electrokinetic remediation for multi-metal contaminated paddy soil using two-dimensional electrode configuration. *Environ. Sci. Pollut. Res.* **21**, 4482–4491 (2014).
17. Tarekgni, M. M., Hiruy, A. M. & Dekebo, A. H. Nano zero valent iron (nZVI) particles for the removal of heavy metals (Cd 2+, Cu 2+ and Pb 2+) from aqueous solutions. *RSC Adv.* **11**, 18539–18551 (2021).
18. Marshall, T., Gravelle, A., Marangoni, A. G., Elsayed, A. & Pensini, E. Zein for hydrocarbon remediation: Emulsifier, trapping agent, or both? *Colloids Surf. A* **589**, 124456 (2020).
19. Tsang, D. C., Zhang, W. & Lo, I. M. Copper extraction effectiveness and soil dissolution issues of EDTA-flushing of artificially contaminated soils. *Chemosphere* **68**, 234–243 (2007).
20. Pensini, E., Laredo, T., Earned, L., Marangoni, A. G. & Ghazani, S. M. A three in onecomplexing agent enables copper desorption from polluted soil, its removal from groundwater and its detection. *Colloids Surf. A* **624**, 126840 (2021).
21. Xu, Y. et al. Simultaneous removal of benzene, toluene, ethylbenzene and xylene (BTEX) by CaO2 based Fenton system: Enhanced degradation by chelating agents. *Chem. Eng. J.* **331**, 255–264 (2018).
22. Earned, L., Laredo, T., Marangoni, A. G. & Pensini, E. Fenton’s degradation of toluene using chelating and emulsifying surfactants. *Int. J. Environ. Sci. Technol.* **1**, 1–14 (2021).
23. Marshall, T., Marangoni, A. G., Lim, L. T., Tchoukow, P. & Pensini, E. Oxidizing emulsifiers: Gelators for water in hydrocarbon reactive emulsions. *J. Environ. Chem. Eng.* **9**, 104998 (2020).
24. Liang, C. & Chen, C.-Y. Characterization of a sodium persulfate sustained release rod for in situ chemical oxidation groundwater remediation. *Ind. Eng. Chem. Res.* **56**, 5271–5276 (2017).
25. Mosheri, H., Alaie, E., Shavandi, M., Dastghheib, S. M. M. & Tasharrofi, S. Benzene-contaminated groundwater remediation using calcium peroxide nanoparticles: Synthesis and process optimization. *Environ. Monit. Assess.* **189**, 452 (2017).
26. Minetti, R. C. P., Macaño, H. R., Britch, J. & Allende, M. C. In situ chemical oxidation of BTEX and MTBE by ferrate: pH dependence and stability. *J. Hazard. Mater.* **324**, 448–456 (2017).
27. Leu, J., O’Connell, S. & Bettahar, M. Remedial process optimization and ozone sparging for petroleum hydrocarbon-impacted groundwater. *Remediat. J.* **26**, 73–94 (2016).
28. Shores, A. R., Hethcok, B. & Latturi, M. Phytoremediation of BTEX and Naphthalene from produced-water spill sites using Poaceae. *Int. J. Phytorem.* **20**, 823–830 (2018).
29. Krouzek, J. et al. Pilot scale applications of microwave heating for soil remediation. *Chem. Eng. Process. Process Intensif.* **130**, 53–60 (2018).
30. Stefanakis, A. L., Seeger, E., Dorer, C., Sinke, A. & Thullner, M. Performance of pilot-scale horizontal subsurface flow constructed wetlands treating groundwater contaminated with phenols and petroleum derivatives. *Ecol. Eng.* **95**, 514–526 (2016).
31. Jousse, F., Atteia, O., Höhener, P. & Cohen, G. Removal of NAPL from columns by oxidation, sparging, surfactant and thermal treatment. *Chemosphere* **188**, 182–189 (2017).
32. Pensini, E., Tchoukow, P., Yang, F. & Xu, Z. Effect of humic acids on bitumen films at the oil-water interface and on emulsion stability: Potential implications for groundwater remediation. *Colloids Surf. A* **544**, 53–59 (2018).
33. Dahal, G., Holcomb, J. & Socci, D. Surfactant-oxygen co-application for soil and groundwater remediation. *Remediat. J.* **26**, 101–108 (2016).
34. Gupta, P.K., & Yadav, B.K. Bioremediation of non-aqueous phase liquids (NAPLs) polluted soil and water resources, in: R.N. Bharagava (Ed) Environmental Pollutants and their Bioremediation Approaches CRC Press, Boca Raton, 2017, pp. 241–256.
35. Khodaei, K., Nassery, H. R., Asad, M. M., Mohammadzadeh, H. & Mahmoodlu, M. G. BTEX biodegradation in contaminated groundwater using a novel strain (Pseudomonas sp. BTEX-30). *Int. Biodeterior. Biodegradation* **116**, 234–242 (2017).

https://doi.org/10.1038/s41598-022-20241-4
36. Switzer, C., Pironi, P., Gerhard, J. I., Rein, G. & Torero, J. I. Volumetric scale-up of smouldering remediation of contaminated materials. *J. Hazard. Mater.* **268**, 51–60 (2014).

37. Vidonish, J. E., Zygonkuras, K., Masiello, C. A., Sabadell, G. & Alvarez, P. J. Thermal treatment of hydrocarbon-impacted soils: A review of technology innovation for sustainable remediation. *Engineering* **2**, 426–437 (2016).

38. Earnard, L., Marangoni, A. G., Gregori, S., Paschos, A. & Pensini, E. Zein-based graphene and biosurfactants enable the electrokinetic clean-up of hydrocarbons. *Langmuir* **37**, 11153–11169 (2021).

39. Maturi, K., Reddy, K. R. & Cameselle, C. Sulfactant-enhanced electrokinetic remediation of mixed contamination in low permeability soil. *Sep. Sci. Technol.* **44**, 2385–2409 (2009).

40. Lo, I. M. & Hu, L. M. Long-term migration of light nonaqueous-phase liquids in two unsaturated soils: Clayey silt and fine sand. *Pract. Period. Hazard. Toxic Radioact. Waste Manag.* **8**, 228–237 (2004).

41. Powers, S. E. et al. The transport and fate of ethanol and BTEx in groundwater contaminated by gasohol. *Crit. Rev. Environ. Sci. Technol.* **31**, 79–123 (2001).

42. Larter, S. et al. An experimental investigation of geochromatography during secondary migration of petroleum performed under subsurface conditions with a real rock. *Geochim. Trans.* **1**, 54 (2000).

43. Switzer, C., Pironi, P., Gerhard, J. I., Rein, G. & Torero, J. L. Volumetric scale-up of smouldering remediation of contaminated soil. *Langmuir* **37**, 11153–11169 (2021).

44. Marshall, T. et al. Path-dependent rheology of carbon particle-hydroxycellulose fluids. *Colloids Surf., A* **612**, 126600 (2020).

45. Guerin, T. E., Horner, S., McGovern, T. & Davey, B. An application of permeable reactive barrier technology to petroleum hydrocarbon contaminated groundwater. *Water Res.* **36**, 15–24 (2002).

46. Vignola, R. et al. Zeolites in a permeable reactive barrier (PRB): One year of field experience in a refinery groundwater—Part I: The performances. *Chem. Eng. J.* **178**, 204–209 (2011).

47. Zhao, Q., Choo, H., Bhatt, A., Burns, S. E. & Bate, B. Review of the fundamental geochemical and physical behaviors of organo-clays in barrier applications. *Appl. Clay Sci.* **142**, 2–20 (2017).

48. Do, S.-H., Kwon, Y.-J. & Kong, S.-H. Feasibility study on an oxidant-injected permeable reactive barrier to treat BTEx contamination: Adsorptive and catalytic characteristics of waste-reclaimed adsorbent. *J. Hazard. Mater.* **191**, 19–25 (2011).

49. Liang, S.-H., Kao, C.-M., Kuo, Y.-C. & Chen, K.-F. Application of persulfate-releasing barrier to remediate MTBE and benzene contaminated groundwater. *J. Hazard. Mater.* **185**, 1162–1168 (2011).

50. Lin, C.-W., Chen, L.-H., Yeh-Pole, I. & Lai, C.-Y. Microbial communities and biodegradation in lab-scale BTEx-contaminated groundwater remediation using an oxygen-releasing reactive barrier. *Bioprocess Biosyst. Eng.* **33**, 383–391 (2010).

51. Marshall, T. et al. Selective solvent filters for non-aqueous phase liquid separation from water. *Sci. Rep.* **10**, 1–13 (2020).

52. Hu, D. et al. Performance and extracellular products analysis of a pilot scale anaerobic membrane bioreactor for treating tetrahydrofuran pharmaceutical wastewater at different HRTs. *J. Hazard. Mater.* **342**, 383–391 (2018).

53. Meix, X. et al. Acetonitrile wastewater treatment enhanced by a hybrid membrane- aerated bioreactor containing aerated and non-aerated zones. *Biore. Technol.* **289**, 121754 (2019).

54. Li, T., Liu, J., Bai, R., Onda, D. G. & Wong, F. S. Biodegradation of organonitriles by adapted sludge consortium with acetonitrile-degrading microorganisms. *Water Res.* **41**, 3465–3473 (2007).

55. Raj, C. C. & Quen, H. L. Advanced oxidation processes for wastewater treatment: Optimization of UV/H2O2 process through a statistical technique. *Chem. Eng. Sci.* **60**, 5305–5311 (2005).

56. Wang, Y. et al. Green recovery of hazardous acetonitrile from high-salt chemical wastewater by pervaporation. *J. Clean. Prod.* **197**, 742–749 (2018).

57. Mei, X. et al. A novel system for zero-discharge treatment of high-salinity acetonitrile-containing wastewater: Combination of pervaporation with a membrane-aerated bioreactor. *Chem. Eng. J.* **384**, 123338 (2020).

58. Longers, O., Gauthier, A., Ballerat-Bussoroles, K. & Andanson, J. M. Deep eutectic solvent with thermo-switchable hydrophobicity. *ACS Sustain. Chem. Eng.* **8**, 12516–12520 (2020).

59. Tabata, M., Kumamoto, M. & Nishimoto, J. Chemical properties of water-miscible solvents separated by salting-out and their application to solvent extraction. *Anal. Sci.* **10**, 383–388 (1994).

60. Souza, R. L., Lima, R. A., Coutinho, J. A., Soares, C. M. & Lima, A. S. Aqueous two-phase systems based on chloroform salts and tetrahydrofuran and their use for lipase purification. *Sep. Purif. Technol.* **155**, 118–126 (2015).

61. Wang, B., Ezejas, T., Feng, H. & Blaschek, H. Sugaring-out: A novel phase separation and extraction system. *Chem. Eng. Sci.* **63**, 2595–2600 (2008).

62. Dhamole, P. B., Mahajan, P. & Feng, H. Phase separation conditions for sugaring-out in acetonitrile–water systems. *J. Chem. Eng. Data* **55**, 3803–3806 (2010).

63. de Brito Cardoso, G. et al. Novel aqueous two-phase systems composed of acetonitrile and polyols: phase diagrams and extractive performance. *Sep. Purif. Technol.* **124**, 54–60 (2014).

64. Marshall, T., Earnard, L., Marangoni, A. G., Laredo, T. & Pensini, E. Cubic mesophases of self-assembled amphiphiles separate miscible solvents. *Colloids Surf., A* **650**, 129548 (2022).

65. Suielli, G. & Bagno, A. Preferential solution of glucose and talose in water–acetonitrile mixtures: A molecular dynamics simulation study. *Phys. Chem. Chem. Phys.* **12**, 2981–2988 (2010).

66. Suzuki, T. The hydration of glucose: The local configurations in sugar–water hydrogen bonds. *Phys. Chem. Chem. Phys.* **10**, 96–105 (2008).

67. Andrić, J. M., Janičić, G. V., Ninković, D. B. & Zarić, S. D. The influence of water molecule coordination to a metal ion on water hydrogen bonds. *Phys. Chem. Chem. Phys.* **14**, 10896–10898 (2012).

68. Chowdhuri, S. & Chandra, A. Dynamics of halide ions: Water hydrogen bonds in aqueous solutions: Dependence on ion size and temperature. *J. Phys. Chem. B* **110**, 9674–9680 (2006).

69. Politi, R., Sapir, L. & Harries, D. The impact of polyls on water structure in solution: A computational study. *J. Phys. Chem. A* **113**, 7548–7555 (2009).

70. Nemati, M., Mogaddam, M. A., Farazajdeh, M. A., Tuzen, M. & Khandaghi, J. In-situ formation/decomposition of deep eutectic solvent during solidification of floating organic droplet-liquid microextraction method for the extraction of some antibiotics from honey prior to high performance liquid chromatography-tandem mass spectrometry. *J. Chromatogr. A* **1660**, 462653 (2021).

71. Nemati, M., Tuzen, M., Farazajdeh, M. A., Kaya, S. & Mogaddam, M. R. A. Development of dispersive solid-liquid extraction method based on organic polymers followed by deep eutectic solvents elution; application in extraction of some pesticides from milk samples prior to their determination by HPLC-MS/MS. *Anal. Chim. Acta* **33**, 9570 (2022).

72. Jouban, A., Farazajdeh, M. A., Nemati, M., Nabil, A. A. A. & Mogaddam, M. R. A. Preparation of ferrofluid from toner powder and deep eutectic solvent used in air-assisted liquid-liquid microextraction: Application in analysis of sixteen polycyclic aromatic hydrocarbons in urine and saliva samples of tobacco smokers. *Microchem. J.* **154**, 104631 (2020).

73. Diaz, R. M., Fernandez, A. M. & Xiberta, J. Extraction of copper from aqueous solutions with carboxylic acids. *Chem. Tech. Biotechnol.* **56**, 273–280 (1986).

74. Chang, J., Yoo, S., Lee, W., Kim, D. & Kang, T. Spontaneous phase transfer-mediated selective removal of heavy metal ions using biocompatible oleic acid. *Sci. Rep.* **7**, 1–7 (2017).
75. L. Eardten, T. Laredo, A.G. Marangoni, S. Mirezae Ghazani, E. Pensini, Modulation of the Viscosity of Guar-Based Fracking Fluids Using Salts, Energy & Fuels, 35 (2021) 16007–16019.

76. Marshall, T., Marangoni, A., Laredo, T., Al-Abdul-Wahid, M.S. & Pensini, E. Mechanisms of solvent separation using sugars and sugar alcohols. Colloids Surf. A 642, 128707 (2022).

77. Levenberg, K. A method for the solution of certain non-linear problems in least squares. Q. Appl. Math. 2, 164–168 (1944).

78. Marquardt, D. W. An algorithm for least-squares estimation of nonlinear parameters. J. Soc. Ind. Appl. Math. 11, 431–441 (1963).

79. Max, J. I. & Chapados, C. Sucrose hydrates in aqueous solution by IR spectroscopy. J. Phys. Chem. A 105, 10681–10688 (2001).

80. Lenz, A. & Ojamäe, L. Theoretical IR spectra for water clusters (H2O)n (n= 6–22, 28, 30) and identification of spectral contributions from different H-bond combinations in gaseous and liquid water. J. Phys. Chem. A 110, 13388–13393 (2006).

81. Marshall, T., Paschos, A., Marangoni, A.G., Yang, F. & Pensini, E. Injectable cationic tramps and sticky bacterial emulsifiers: A safe alliance during diesel bioremediation. Colloids Surf. A 613, 126051 (2020).

82. Leontowich, A. F. et al. The lower energy diffusion and scattering side-bounce beamline for materials science at the Canadian Light Source. J. Synch. Radiat. 28, 1 (2021).

83. Von Dreelle, R. B. Small-angle scattering data analysis in GSAS-II. J. Appl. Crystalllogr. 47, 1784–1789 (2014).

84. Toby, B. H. & Von Dreelle, R. B. GSAS-II: the genesis of a modern open-source all purpose crystallography software package. J. Appl. Crystalllogr. 46, 544–549 (2013).

85. Teubner, M. & Strey, R. Origin of the scattering peak in microemulsions. J. Chem. Phys. 87, 3195–3200 (1987).

86. Egorov, G. I. & Makarov, D. M. Densities and thermal expansions of (water+ tetrahydrofuran) mixtures within the temperature range from (274.15 to 333.15) K at atmospheric pressure. J. Mol. Liquids 310, 113105 (2020).

87. Pan, Y., Tikekar, R. V., Wang, M. S., Avena-Bustillos, R. J. & Nitin, N. Effect of barrier properties of zein colloidal particles and oil-in-water emulsions on oxidative stability of encapsulated bioactive compounds. Food Hydrocolloids 43, 82–90 (2015).

88. Das, B., Roy, M. N. & Hazra, D. K. Densities and viscosities of the binary aqueous mixtures of tetrahydrofuran and 1, 2-dimethoxyethane at 298, 308 and 318 K. Indian J. Chem. Technol. (IJCT) 1, 93–97 (1994).

89. Critchfield, F. E., Gibson, J. A. Jr. & Hall, J. L. Dielectric constant and refractive index from 20 to 35° and density at 25° for the system tetrahydrofuran—Water. J. Am. Chem. Soc. 75, 6044–6045 (1953).

90. Niu, B. et al. Effect of pH on formation of starch complexes with lauric acid and β-lactoglobulin 132, 109915 (2020).

91. Morrow, B. H., Koenig, P. H. & Shen, J. K. Atomistic simulations of pH-dependent self-assembly of micelle and bilayer from fatty acids. J. Chem. Phys. 137, 194902 (2012).

92. Wang, D. & Jingcheng, H. Self-assembly fibrillar network gels of simple surfactants in organic solvents. Langmuir 27, 1713–1717 (2011).

93. Niraula, B. B., Seng, T. N. & Misran, M. Vesicles in fatty acid salt–fatty acid stabilized o/w emulsion—emulsion structure and rheology. Colloids Surf. A 236, 7–22 (2004).

94. Saleeb, F. Z., Cante, C. J., Streckfus, T. K., Frost, J. R. & Rosano, H. L. Surface pH and stability of oil-water emulsions derived from laureate solutions. J. Am. Oil Chem. Soc. 52, 208–212 (1975).

95. Beaud, P. et al. Reverse micellar extraction of dyes based on fatty acids and recoverable organic solvents. Sep. Purif. Technol. 242, 116772 (2020).

96. Reverse micelles: biological and technological relevance of amphiphilic structures in apolar media (Springer New York, NY, 1984).

97. Shultz, M. J. & Vu, T. H. Hydrogen bonding between water and tetrahydrofuran relevant to clathrate formation. J. Phys. Chem. B 119, 9167–9172 (2015).

98. Wang, S. H. & Griffiths, P. R. Resolution enhancement of diffuse reflectance ir spectra of coals by Fourier self-deconvolution: 1 CH stretching and bending modes. Fuel 64, 229–236 (1985).

99. Cui, Y., Rushing, J. C., Seifert, S., Bedford, N. M. & Kuroda, D. G. Moleularly heterogeneous structure of a nonionic deep eutectic solvent composed of N-methylacetamide and lauric acid. J. Phys. Chem. B 123, 3984–3993 (2019).

100. Van Thiel, M., Becker, E. D. & Pimentel, G. C. Infrared studies of hydrogen bonding of water by the matrix isolation technique. J. Chem. Phys. 27, 486–490 (1957).

101. Francis, S. A. Absolute intensities of characteristic infra-red absorption bands of aliphatic hydrocarbons. J. Chem. Phys. 18, 861–865 (1950).

102. Bailey, S., Froment, G. F. & Waugh, K. C. A DRIFTS study of the morphology and surface adsorbate composition of an operating methanol synthesis catalyst. Catal. Lett. 30, 99–111 (1994).

103. Wu, K., Feng, S., Hedoux, A. & Shalaev, E. Water structure in glycerol: Spectroscopic and computer simulation investigation of hydrogen bonding and water clustering. J. Mol. Liquids 355, 118916 (2022).

104. Purkayastha, D. D. & Madhurima, V. Interactions in water–THF binary mixture by contact angle, FTIR and dielectric studies. J. Mol. Liquids 187, 54–57 (2022).

105. Amine, C., Kreher, H., Helgason, T. & Tadros, T. Investigation of emulsifying properties and emulsion stability of plant and milk proteins using interfacial tension and interfacial elasticity. Food Hydrocolloids, 113105 (2020).

106. Pensini, E. et al. Emulsification mechanism of asphaltenene-stabilized water-in-oil emulsions by a polymeric ethylene oxide–propylene oxide demulsifier. Energy Fuels 28, 6760–6771 (2014).

107. Adamson, D. T., Anderson, R. H., Mahendra, S. & Newell, C. J. Evidence of 1, 4-dioxane attenuation at groundwater sites contaminated with chlorinated solvents and 1, 4-dioxane. Environ. Sci. Technol. 49, 6510–6518 (2015).

108. Isaacson, C., Mohr, T. K. & Field, J. A. Quantitative determination of 1, 4-dioxane and tetrahydrofuran in groundwater by solid phase extraction GC/MS/MS. Environ. Sci. Technol. 40, 7305–7311 (2006).

109. Lindoy, L. F. & Baldwin, D. S. Ligand design for selective metal-ion transport through liquid membranes. Pure Appl. Chem. 61, 909–914 (1989).

110. Yoshida, Y. et al. Ion transfer reactions across the aqueous|organic solution interface in the presence of phospholipid layer adsorbed on the interface i1037–i1039 (The Japan Society for Analytical Chemistry, 2001).

111. Wojciechowski, K., Kucharek, M. & Buille, J. Mechanism of Cu(III) transport through permeation liquid membranes using Azacrown ether and fatty acid as carrier. J. Membr. Sci. 314, 152–162 (2008).

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