**INTRODUCTION**

Plastic pollution is a problem of growing environmental concern, because production rates have increased to 380 million metric tons (MT) year^{-1} in 2015 (1), and annual waste production is projected to rise to 3.4 billion MT in the next 30 years (2). Many countries still have inefficient waste management and water treatment systems allowing leakage to the environment (3), which is exacerbated by littering behavior. Mismanaged plastic waste could triple from 60 million to 99 million MT in 2015 to 155 to 265 million MT by 2060, assuming a business-as-usual scenario (4). In addition, plastic is designed to be durable. Therefore, it persists in the environment for long periods of time. Thus, it is hardly unexpected that plastic pollutants are ubiquitous and have been reported from environments close to urban centers, terrestrial areas, and freshwater environments, as well as from the shores of remote uninhabited islands, the sea surface, water column, and deep seafloor (5). Plastic pollutants have also reached polar regions including Arctic beaches (6), sea ice (7), water column (8), sea surface (9–12), and the seafloor (13). Under the influence of light, mechanic abrasion, waves, and temperature fluctuations, plastic items fragment into smaller particles and are termed microplastic (MP) when attaining sizes below 5 mm. Unexpectedly, Arctic surface waters turned out to harbor the highest MP concentrations in a global comparison of MP quantities, despite their remoteness (14). In addition, litter quantities have increased significantly on the deep Arctic seafloor over the past 15 years as highlighted in a time-series study (13). This stimulated further research unveiling very high concentrations of MPs in the sediment of the deep Fram Strait (15) and Arctic sea ice (16), prompting the question: “How do all these MPs make it so far to the north?”

One possibility is aerial transport since a yet still limited number of studies found MPs in atmospheric fallout of the cities Dongguan (China) and Tehran (17, 18). In France, MP concentrations in atmospheric fallout increased fivefold after a rain event, indicating that wet deposition could be a pathway of MPs to Earth’s surfaces (19) including the oceans. Despite the limited research on airborne plastics, its importance is obvious given the precautions that have to be taken in studies on MPs to reduce the risk of contamination through airborne MPs (20). Further, airborne MPs represent a hitherto largely neglected route of exposure to humans and wildlife as these particles could be taken up through inhalation (21). Snow is a scavenger for diverse impurities and acts as a filter on the ground by dry deposition (22, 23). Here, we sampled snow deposited on ice floes drifting in Fram Strait and on Svalbard to assess whether atmospheric fallout is a pathway of MPs to the Arctic environment. To enable comparisons with previous studies in the Arctic, harmonized automated Fourier transform infrared (FTIR) imaging analysis was applied (15, 16, 24, 25). The Arctic is still widely conceived as one of the last pristine environments on the globe. To verify this view, we also assessed MP concentrations in snow close to more urban sites in northern Europe (Bremen City, Isle of Heligoland) and the Alps (Davos, Tschuggen, Bavaria) for comparison (Fig. 1).

**RESULTS**

MPs and microfibers were found in all but one snow samples, ranging from 0.02 × 10^3 to 154 × 10^3 N liter^{-1} and 0.043 × 10^3 to 10.2 × 10^3 N liter^{-1}, respectively (Fig. 2 and table S1). MPs accounted for 0 to 88% (mean, 12%) of the filtered particles. The material type of fibers, however, could not be determined because of the applied analysis pipeline. Therefore, fiber numbers comprise both synthetic and natural fibers. Nevertheless, we present fiber concentrations to enable comparison with previously published studies, which reported primarily fibers.

**Fiber and MP quantities at different locations**

MPs occurred at a mean concentration of 9.8 × 10^3 N liter^{-1} (± SE 6.9 × 10^3 N liter^{-1}) with highest quantities detected in the sample Bavaria 2 (154 × 10^3 N liter^{-1}), followed by Heligoland 2 (17.6 × 10^3 N liter^{-1}) and Ice Floe 9 (14.4 × 10^3 N liter^{-1}) (Fig. 1, A and B).
and table S1). Snow from Europe contained significantly more MPs ($24.6 \pm 18.6 \times 10^3 \text{ N liter}^{-1}$) than samples from ice floes (Mann-Whitney U test: $W = 54, P = 0.011$). Still, a mean of $1.76 \pm 1.58 \times 10^3 \text{ N liter}^{-1}$ is substantial for a secluded location such as the Arctic. Figure 3 illustrates images of MPs and fibers from different locations. Ice Floe 9 had a much higher MP concentration than the other Arctic snow samples, raising concern regarding the potential of contamination during sampling or from the helicopter used. Exclusion of this value decreases the mean to $0.18 \pm 0.07 \times 10^3 \text{ N liter}^{-1}$. Still, we think that this high level is unlikely due to contamination as varnish, the main polymer type detected, was not part of the equipment used and also present in high quantities in samples from Ice Floe 3, Bremen, the Alps, and Heligoland, where no helicopter was used. In addition, the sample Ice Floe 9 was taken outside the area likely to be affected by helicopter turbulence. Since the second highest MP abundance was detected on Heligoland at a time of strong winds ($18 \text{ m s}^{-1}$; table S2), we tested whether MP abundance was correlated with wind speed but found no significant correlation (Spearman’s rank correlation test: $\rho = -0.06, P = 0.80$). However, wind direction may have played a role as the strong easterly winds may have transported particles from the neighboring dunes or from the mainland.

The maximum quantities of fibers were detected in snow from Ice Floe 4 ($10.2 \times 10^3 \text{ N liter}^{-1}$), followed by Heligoland 2 ($2.57 \times 10^3 \text{ N liter}^{-1}$) and Bavaria 3 ($2.75 \times 10^3 \text{ N liter}^{-1}$) (Fig. 1, A and B, and table S1). Still, overall, snow from Europe contained significantly more fibers (mean, $1.431 \pm 0.325 \times 10^3 \text{ N liter}^{-1}$) than samples from Arctic ice floes (Mann-Whitney U test: $W = 96, P = 0.024$), which had still a substantial mean ($1.38 \pm 1.10 \times 10^3 \text{ N liter}^{-1}$) given their remoteness. Again, it could be argued that the exceptionally high fiber numbers from Ice Floe 4 may be due to contamination. Its exclusion would lead to a lower mean ($0.28 \pm 0.095 \times 10^3 \text{ N liter}^{-1}$). However, we deem this unlikely as Ice Floe samples 5 to 8 were taken on the same large ice floe by the same personnel, so contamination of these samples should be similarly high. There was no significant correlation between fiber abundance and wind speed (Spearman’s rank correlation test: $\rho = -0.17, P = 0.45$). The abundance of fibers was positively correlated with MP abundance (Spearman’s rank correlation test: $\rho = 0.62, P = 0.002$).

**Size of MPs and fibers**

The size of MPs detected ranged between 11 and 475 μm. Eighty percent of all detected MPs were $\leq 25 \mu m$, and 98% of all particles were $< 100 \mu m$. Overall, the amount of particles decreased with increasing size (Fig. 2C) with no asymptote reached in the smallest detectable size class. Table S2 comprises the size distribution of individual samples.
The length of fibers ranged between 65 and 14,314 μm. While 97% had a maximum length of 5 mm, 31% were shorter than 500 μm. In general, the fibers show an increasing trend toward shorter lengths but are not saturated in the lowest size class (fig. S1). European fibers were significantly longer compared with those from Arctic snow (Mann-Whitney U test: $W = 13,723$, $P = 0.0001$) even if the two largest European fibers (14,314 and 13,704 μm) were excluded. Table S3 includes the size information of individual fibers marked with their region of origin.

**Material composition**
The highest proportion of MPs in the total natural and synthetic particle load was found in snow from Ice Floe 1 (88%), followed by Bavaria 2 (67%) and Ice Floe 9 (37%) (table S2). There was no significant difference in the proportion of MP particles from European and Arctic snow (Mann-Whitney U test: $W = 170$, $P = 0.59$). The composition varied considerably with 19 different polymer types found in total ranging between 2 (Ice Floe 4) and 12 types (Bavaria 2) per sample (Fig. 2D and table S1). The number of polymers per sample was significantly higher in European (mean, $8.63 \pm 0.80$) compared with Arctic (mean, $5.14 \pm 0.79$) samples (Mann-Whitney U test: $W = 123$, $P = 0.013$). Acrylates/polyurethanes/varnish/lacquer (hereafter varnish) occurred most frequently (17 samples), followed by nitrile rubber (16 samples), polyethylene (PE), polyamide, and rubber type 3 (13; ethylene-propylene-diene rubber). The polymer composition of samples from Europe...
and the Arctic was significantly different \([\text{permutational multivariate analysis of variance (PERMANOVA)}]: \text{pseudo}-F = 2.43, P = 0.006\]. The dissimilarity in the polymer composition from European and Arctic samples was 67% and caused primarily by much higher abundances of polyamide, varnish, rubber type 3, nitrile rubber, ethylene-vinyl-acetate, and PE in European samples. By contrast, polystyrene, polyvinyl chloride (PVC), polycarbonate, polylactic acid, and polyimide occurred exclusively in Arctic snow.

**Other particles**

Other particles detected in snow accounted for 22 to 100% of the total particles (table S2) and comprised chitin, charcoal, coal, animal fur, plant fibers, and sand. Except for coal, all of these were significantly more abundant in European snow (Mann–Whitney \(U\) test: \(W = 114, P = 0.0015\)), which also explains the significant differences found in their composition (PERMANOVA: pseudo-\(F = 7.75, P = 0.001\)). Particles assigned to “plant fibers” and “animal fur” contributed most to the 46% dissimilarity [similarity percentage (SIMPER)].

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**Fig. 3. Photographs of MPs detected in snow.** (A) Polystyrene fiber from Svalbard 4 (length, 1101 \(\mu m\)); (B) polypropylene particle from Heligoland (diameter, 256 \(\mu m\)); (C) polyvinyl chloride fiber from Ice Floe 8 (length, 956 \(\mu m\)). Note that these particular two fibers could be analyzed by FTIR as they happened to lay plane on the filter. (D to F) Aluminum oxide filter with enriched snow sample. (G to I) Corresponding polymer-dependent false-color image after FTIR measurement and automated analysis. (D and G) Bavaria 2: The sample with most MPs; (E and H) Ice Floe 9: the sample with the third most MPs; (F and I) Bremen: The sample with intermediate MP numbers but many fibers (photo credit: S. Mützel, Alfred-Wegener-Institut Helmholtz-Zentrum für Polar- und Meeresforschung).
DISCUSSION

This study provides the first data on contamination of snow by MPs. MP concentrations in snow were very high, indicating significant contamination of the atmosphere. During its passage through the atmosphere, snow binds airborne particles and pollutants, which are eventually deposited on Earth’s surfaces, a phenomenon termed “scavenging” (22). Our data show that scavenging represents an important pathway of MPs to land and ocean environments in Europe and the Arctic. On the basis of annual snow fall data, we estimate an annual MP deposition of 8.8 ± 7.9 N m⁻² (0 to 72 N m⁻²) in the Fram Strait, 1.4 ± 0.4 N m⁻² in Svalbard (0 to 2 N m⁻²), and 66 ± 60.1 N m⁻² (0 to 308 N m⁻²) in the Alps. It should be noted, however, that these estimates come with very large uncertainties given the variability of the data.

Abundance of MP particles and microfibers

Although MP levels were significantly higher in European (0.191 × 10³ to 154 × 10³ N liter⁻¹) compared with Arctic snow (0 to 14.4 × 10³ N liter⁻¹), the conclusions drawn have to be treated with caution given the high variability and sparse spread of samples over large areas. Some variability may be due to the fact that snow from the Arctic was not freshly deposited unlike all European snow samples (except for Swiss snow) and could have been exposed to secondary dry deposition of airborne particles (23) for an unknown period of time. Still, the lower concentrations found in Arctic snow is not unusual given the distance of the Arctic to densely populated source areas. MP concentrations were also consistently higher in atmospheric fallout from densely populated urban compared with suburban or less densely populated sites in France and Dongguan (18, 26). Dehghani et al. (17) recorded MPs in the range of 3 to 20 N per gram of dry dust from Tehran. However, quantities are not directly comparable as wet deposition is a more efficient mode of transport than dry deposition (27), and different methods were used. The same applies to MPs recorded in dry fallout from Dongguan (31 to 43 N m⁻² day⁻¹) (18).

The highest MP concentration was found in Bavarian snow. Since we took this sample next to a country road, traffic could play a role in terms of automotive emissions, dispersion of settled MPs by cars, and subsequent scavenging by falling snow. Experiments have shown that significant numbers of similar sized zinc sulfide particles placed on a road were swirled by cars (28). While the most abundant polymer types in this sample, rubber types 1 and 3, point to car tires as a potential source, no firm conclusions can be drawn because of current methodological constraints in the detection of specific rubber types. Still, the number of MPs in snow from a street in the city of Bremen, our most urban site (~568,000 inhabitants), was only intermediate (2 × 10³ N liter⁻¹). This could be due to lower particle dispersion as our most urban site (~568,000 inhabitants), was only intermediate still, the number of MPs in snow from a street in the city of Bremen, methodological constraints in the detection of specific rubber types. While the most abundant polymer on a road were swirled by cars (28). While the most abundant polymer types in this sample, rubber types 1 and 3, point to car tires as a potential source, no firm conclusions can be drawn because of current methodological constraints in the detection of specific rubber types.

Comparison with MP levels in Arctic sea ice and deep-sea sediments

Unfortunately, a direct comparison of our results on MP particles with previous data on airborne MPs or those from Arctic seawater is not possible as the latter two deal primarily with fibers (but see the “Abundance of fibers” section) or use different methods resulting in much lower numbers (9, 10, 35). We can, however, compare our results of Arctic ice floes with previous data on MPs in Arctic sea ice and deep-sea sediments (15, 16), also to investigate whether airborne MPs are a source to these compartments. These studies used the same harmonized analytical methods and are thus comparable with our approach. MP concentrations ranged between 1.1 × 10³ and 12 × 10³ N liter⁻¹ in Arctic sea ice, which tends to concentrate particles by a factor of 100 relative to adjacent seawater (16) and between 0.04 × 10³ and 3.46 × 10³ N liter⁻¹ in deep-sea sediments (15). Given similar or higher levels in snow from ice floes (0 to 14.4 × 10³ N liter⁻¹), it seems likely that they contribute to MPs in
sea ice and also sink to the seafloor over time. MPs deposited on ice floes via snow may be embedded directly into the sea ice matrix through snow metamorphism, i.e., compaction, or through release into melt ponds and subsequent refreezing. Another route to these matrices could simply be through wet deposition onto the ocean surface before or during sea ice formation. Once in seawater, MPs may sink to the seafloor. The composition of polymers in snow from ice floes seemed more similar to that of sea ice compared with deep-sea sediments. For example, varnish was one of the most frequent and abundant polymer types reported in both of these spheres but less important in sediments. Still, nitrile rubber and polyamide were important in both ice floe snow and deep-sea sediments and sea ice, such that these polymers may have been transported from the atmosphere to sea ice and the seafloor.

**MP size**

The size distribution of MPs in snow was unexpectedly similar to MP sizes found in Arctic sea ice and deep-sea sediments (15, 16). Most of the MPs detected in all of these spheres were in the smallest size range (11 μm). Scrutinized by scanning electron microscopy, MP particles from atmospheric fallout of Dongguan revealed signs of weathering such that Cai et al. (18) hypothesized atmospheric degradation processes through collision and friction dynamics, as well as chemical weathering due to higher irradiation and oxygen levels in the atmosphere. This could enhance fragmentation into smaller sizes. The fact that there is no size saturation in the smallest size range implies that there may be yet more particles in size categories below our current detection limit. In addition, smaller particles are probably more likely to be picked up and transported by air masses. No matter what the cause, our results likely underestimate MP quantities as was also concluded in previous studies (15, 16, 36). This highlights the need to quantify small particles for realistic assessments of MP pollution. Most studies currently focus on particles >200 to 300 μm in size.

**Abundance of fibers**

All snow samples contained fibers ranging between 0.043 × 10^3 and 10.2 × 10^3 N liter⁻¹ with highest concentrations in snow from Ice Floe 4 followed by snow from Bremen. While it is unknown what proportion of these were synthetic polymers, our fiber concentrations were at least four to seven orders of magnitude higher than previous reports from outdoor and indoor environments near Paris (median, 0.0009 and 0.0054 N liter⁻¹, respectively) (36). Fiber and MP abundance in Dongguan were in the same order of magnitude as samples from Paris (18). This large discrepancy is likely due to methodological differences and differences in the underlying mechanisms, i.e., dry versus wet deposition. Dris et al. (19) reported five times higher MP levels after a rain event, and snow can scavenge aerosol particles up to 50 times more efficiently than rain (22). Whatever the reasons, the high abundance of fibers in remote Arctic snow is notable, whereas high levels in urban Bremen do not come as a surprise. The positive correlation between fiber and MP concentrations may indicate that they both come from similar sources. While 33% of the fibers detected near Paris were polymers, 23% of the fibers from Dongguan were made of plastic (18, 36). If 28% (the mean of these two values) of the fibers from our study were also polymers, then this would add 0.012 × 10^3 to 2.86 × 10^3 N liter⁻¹ and result in a total MP load of 0.029 × 10^3 to 156.86 × 10^3 N liter⁻¹. However, this assumption has to be treated with caution since the proportion of natural fibers in our samples may differ from those of more urban areas.

**Polymer composition**

The polymer composition of samples was very variable, even at sites which were located in close proximity, e.g., Ice Floe 4 to Ice Floe 8, that came from the same ice floe. Scavenging of particles depends on the size of snow crystals, particle size, wind speed, air humidity, and snow intensity (22, 37) to name a few of many factors that may have influenced also the polymer composition. Variable polymer compositions were also reported from Arctic sea ice and deep-sea sediments (15, 16). The number of polymer types found per sample was highest in Bavaria (12 types), which is not unusual given that this was also the sample with the highest MP abundance and sampled next to a country road where settled MPs may be subject to dispersion by traffic. The number of detected polymer types both per sample and for all samples was much higher in this study (19) compared with atmospheric fallout from China (4) and France (3) (18, 26). This is likely due to wet deposition scavenging effects and methodological differences, because previous results were based on the analysis of visually preselected particles and considered only particles larger than 50 μm, which were also mostly fibers.

Varnish was the most frequent and among the top five most abundant polymer types. It was also detected in ice cores and deep-sea sediments albeit less frequently (15, 16). This reflects the widespread application of polymer-based varnish for protective coatings of surfaces, including vehicles, ships, wind turbines, aquaculture and buildings. The building sector accounts for ~20% of the European plastic converter demand (38). MPs may be emitted during construction on building sites, through abrasion of coated surfaces by wind and rain or when handling construction waste (39). Future research is needed to quantify the importance and pathway of this source of MPs.

Three different types of rubber were the next most frequent and the most abundant polymers. They may enter the environment as abrasion product from tires or as extremely durable synthetic rubber roofing membranes commonly used in roof construction (rubber type 3, ethylene-propylene-diene rubber). Abrasion from sealing gaskets (rubber type 1, sealing rubber), cable, and shoe soles could be further sources of rubber, some of which belong to the most common sources of MPs in Germany (39). While Arctic sea ice contained scant rubber particles, it was also found in deep-sea sediments (15, 16), where nitrile rubber was one of the most frequent and abundant polymers. As it is very resistant to a wide range of temperatures, as well as to oil, gasoline, and other chemicals, it is widely used for hoses, seals, O-rings, synthetic leather, grommets, cable jacketing, and transmission belts in offshore oil platforms and the automotive and aeronautical industry.

Polyamide (including nylon) was among the top five most frequent and abundant MPs detected. It was also reported from a household and atmospheric fallout in France (36). Polyamide is widely used in synthetic fabrics, automotive applications, fisheries, sails, toothbrushes, packaging, and carpets. It was more prevalent in European than in Arctic snow. Polyamide was, however, abundant in Atlantic and Arctic surface waters and deep-sea sediment (9, 10, 15, 16, 35, 40). Thus, pathways other than atmospheric transport may play a more prominent role in the transport of polyamide to the Arctic.

Likewise, PE occurred only in low numbers in two ice floe samples but was the most abundant polymer type detected in Arctic sea ice and Atlantic surface waters (9, 16, 35, 40) and prevailed in most European samples, as well as in atmospheric fallout of Dongguan (18) and a French household (36). Again, mechanisms other than atmospheric transport may be more important in PE transport to the Arctic.
Health implications

The large concentrations of MPs and microfibers in snow highlight the importance of the atmosphere as a source of airborne MPs and microfibers. Through this pathway, MPs likely find their way into soil and aquatic environments and therefore also into food chains. In populated areas, it is a common practice to remove snow from streets and transport it “away.” Our results show that these locations should be chosen wisely so as to avoid contamination of sensitive areas. Moreover, MPs in snow have fallen out of the atmosphere and could therefore be considered an indicator of aerial MP pollution. This is relevant in the context of human and animal health, especially in terms of inhabited European sites, where residents may breathe in airborne MPs and fibers. However, although we know that airborne contamination of seafood during indoor food preparation and meals exceeds original MP concentrations (41), there has been unexpectedly little research about the inhalation risk of airborne MPs. It has been postulated that only the smallest-sized MP fraction is respired into the deep lung while particles exceeding a length of 5 μm, with a diameter of <3 μm, and with a length-to-diameter ratio of 3:1 are subject to coughing or mucociliary clearance (21) such that they end up in the gastrointestinal tract. Still, the detection of MPs and other fibers of up to 135 μm in length in lung tissues, including carcinoma, challenges this notion (42, 43). MPs in pulmonary tissues may persist for a long time as they are durable in body fluids (21). Chronic inhalation of MPs, especially in combination with adsorbed or added chemicals (17), may lead to health risks including respiratory irritation, allergic alveolitis, inflammation, fibrosis, and genotoxicity (21). Pauly et al. (42) suggest that MPs may be considered candidate agents contributing to the risk of lung cancer, especially of nonsmokers. The high MP concentrations detected in snow samples from continental Europe to the Arctic indicate significant air pollution and stress the urgent need for research on human and animal health effects focusing on airborne MPs.

The high amount of MPs present in the atmosphere as indicated by significant concentrations in snow from continental Europe to the Arctic is puzzling at first. It is known, however, that snow either takes up aerosols or forms around aerosol nuclei containing pollutants, e.g., from vehicle exhaust particles (44). In addition, dust emitted from Earth surfaces, deposition, and dispersion between atmosphere, land surface, and the aquatic realm could facilitate the transportation of MPs (18), but research is needed to verify this. One open question is also whether aquatic environments act as a source of MPs to the atmosphere, e.g., during evaporation and storm events. This could also explain the high concentrations of MPs and fibers in secluded uninhabited regions such as the Arctic. Another explanation is long-range transport by wind systems bringing anthropogenic particles from urban sites. Large dust particles are transported over distances of 3500 km from the Sahara to the North Atlantic (33). This is similar to the distance between our Arctic sites and Europe, which happens to be the most important pathway in terms of wind-driven transport.

Table 1. Details of snow-sampling campaigns. AWI, Alfred-Wegener-Institut. Helmholtz-Zentrum fur Polar- und Meeresforschung; SLF Davos, WSL Institute for Snow and Avalanche Research.

| Sample   | Area     | Longitude °E | Latitude °N | Cruise | Date            | Sampler | Wind speed (m s⁻¹) |
|----------|----------|--------------|-------------|--------|-----------------|---------|-------------------|
| Ice Floe 1 | Arctic   | 79.2576      | 2.3121      | PS99.2 | 07 July 2016    | AWI     | 3.1               |
| Ice Floe 2 | Arctic   | 78.5926      | 5.2376      | PS99.2 | 02 July 2016    | AWI     | 6                 |
| Ice Floe 3 | Arctic   | 79.2576      | 2.3121      | PS99.2 | 30 June 2016    | AWI     | 9–10              |
| Ice Floe 4a | Arctic  | 80.0934      | 0.2253      | PS108  | 30 August 2017  | Aemelire| 2.8–3.6           |
| Ice Floe 5 | Arctic   | 80.0922      | 0.2212      | PS108  | 30 August 2017  | Aemelire| 2.8–3.6           |
| Ice Floe 6 | Arctic   | 80.0915      | 0.2209      | PS108  | 30 August 2017  | Aemelire| 2.8–3.6           |
| Ice Floe 7 | Arctic   | 80.0910      | 0.2163      | PS108  | 30 August 2017  | Aemelire| 2.8–3.6           |
| Ice Floe 8 | Arctic   | 80.0904      | 0.2156      | PS108  | 30 August 2017  | Aemelire| 2.8–3.6           |
| Ice Floe 9 | Arctic   | 79.0315      | 3.5447      | PS107  | 04 August 2017  | AWI     | 2.8–3.6           |
| Svalbard 1 | Arctic   | 78.2118      | 16.5511     | Aemelire| 10 March 2018   | Citizens| 1–3               |
| Svalbard 2 | Arctic   | 77.5107      | 16.2567     | Aemelire| 14 March 2018   | Citizens| 2                 |
| Svalbard 3 | Arctic   | 78.1433      | 15.4215     | Aemelire| 14 March 2018   | Citizens| 6                 |
| Svalbard 4 | Arctic   | 78.2091      | 18.4517     | Aemelire| 16 March 2018   | Citizens| 2                 |
| Svalbard 5 | Arctic   | 77.4586      | 17.0988     | Aemelire| 28 March 2018   | Citizens| 6                 |
| Heligoland 1 | N. Germany | 54.1832 | 7.8884 | - | 17 March 2018 | AWI | 18 |
| Heligoland 2 | N. Germany | 54.1835 | 7.8883 | - | 17 March 2018 | AWI | 18 |
| Bremen   | N. Germany | 53.0675 | 8.7931 | - | 25 February 2018 | AWI | 2.5 |
| Bavaria 1 | Alps     | 47.6505      | 11.4335     | -      | 19 March 2018   | Citizens| 2.5               |
| Bavaria 2 | Alps     | 47.5837      | 11.3921     | -      | 19 March 2018   | Citizens| 2.5               |
| Bavaria 3 | Alps     | 47.4367      | 11.2587     | -      | 19 March 2018   | Citizens| 2.5               |
| Tschuggen | Alps     | 46.7840      | 9.9210      | -      | 07 March 2018   | SLF Davos| 2.5              |
| Davos    | Alps     | 46.7986      | 9.8448      | -      | 07 March 2018   | SLF Davos| 2.5              |
of mercury to the Arctic. Our results highlight a hitherto neglected pathway of MPs to the Arctic Ocean, an ecosystem that is already stressed by the effects of the climate crisis.

MATERIALS AND METHODS

Study sites

Between 2015 and 2017, five ice floes drifting in the Arctic Fram Strait were visited by ship-based helicopters or dinghies during three expeditions of the research icebreaker RV Polarstern (Fig. 1A and Table 1). Surface snow was sampled with a prerinsed mug, a steel spoon, or a soup ladle and transferred into containers made of PVC, PE, or glass (see table S2 for more details). In March 2018, five samples were taken at different locations on Svalbard (Fig. 1A and Table 1) by citizen scientists embarking on a land expedition by ski doo (Aemalire project). The citizens were instructed on contamination prevention and equipped with protocol forms, prerinsed 2-liter stainless steel containers (Ecotanca), a porcelain mug, a steel spoon, and a soup ladle for sampling.

In February 2018, the surface of freshly deposited snow was transferred with a spoon from parking cars in the city of Bremen into glass jars (Fig. 1B and Table 1). Only surface snow was taken, leaving behind a thin layer of snow on top of car surfaces. One month later, freshly deposited surface snow was collected with a soup ladle on the Isle of Heligoland from the backyard and next to a pedestrian path in front of the Alfred Wegener Institute campus (Fig. 1B and Table 1), avoiding soil and bottom surfaces. In the Swiss Alps, surface snow was collected with a spoon at a snowfield in the surroundings of Tschuggen and in the village of Davos (Fig. 1B). This snow had fallen 2 days before sampling (7 March 2018). In the Bavarian Alps (Germany), citizens transferred freshly fallen surface snow with a steel spoon from three different locations into glass jars that had been prerinsed with tap water (Fig. 1B and Table 1): Sample 1 was taken next to a country lane, sample 2 was collected next to a country lane and a lake, and sample 3 was taken in a green area with nearby anthropogenic activity. All teams worked with bare hands, and the equipment was rinsed with Milli-Q water, tap water, or snow. Except for the Bavarian samples, potential sample contamination from the sealing of glass jars was minimized by a sheet of tin foil. Samples were kept frozen but defrosted during transport to Heligoland or in the laboratory.

Contamination prevention and procedural blanks

All personnel taking samples was instructed to position themselves against the wind and sample undisturbed snow in front of them with bare hands to avoid contamination. If not stated otherwise, all laboratory ware was made of glass or stainless steel and thoroughly rinsed with Milli-Q water before use. All polymer-based items, which could not be replaced by alternatives (e.g., bottle caps and filter holders) were made of polytetrafluoroethylene, a polymer that cannot be detected within the current FTIR imaging settings (25). Airborne particles were filtered by dustboxes (DB1000, G4 before filtration, HEPA-H14 final filtration, Q = 950 m³ hour⁻¹, Mocklinghoff Luftertechnik) in laboratories for particle sorting and FTIR analyses. All filtration steps were performed in a laminar flow cabinet (ScanLaf Lufttechnik) in laboratories for particle sorting and FTIR analyses. To account for possible contamination from plastic sample containers, two procedural blanks were created in the laboratory by filling the PVC and PE containers used with Milli-Q water and stored at −20°C for 4 days to simulate the freezing and thawing process. To assess possible contamination of the Bavarian snow samples from used tap water and the resin sealing in lids, a similar glass jar was filled with tap water and frozen by the citizen scientists. All blanks were processed in the same way as the snow samples. For the plastic containers, up to three FTIR imaging runs were performed because of low numbers found to obtain a reliable result, which represents the extent of MP contamination. No blanks were taken during field sampling. The amounts of MPs determined in the samples with available blanks were corrected by converting the quantity of MPs detected in the blank to the sample volumes (table S2). The number of particles per liter was calculated for each sample on the basis of the volume of melted snow. The polymer types, abundance, and sizes of MPs in procedural blanks are available in table S4. Blanks from the PE container harbored 256 N liter⁻¹ and consisted primarily of PE. The PVC flask contained 308 N liter⁻¹, most of which was polypropylene and PVC. The glass jar that had been prerinsed with tap water contained the highest number of MP (27,243 N liter⁻¹) with a diverse mix of polymer types, mostly PE and rubber types 1 and 3. To address the uncertainty of all results regarding possible contamination, the propagation of uncertainty was calculated (45) as described in note S1.

Analytical procedure for the detection of (MP) particles

Following three overhead twists, an aliquot [10 ml: 1 ml of sample (melted snow) diluted with 9 ml of Milli-Q] of each sample was analyzed by FlowCam (Fluid Imaging Technologies, Scarborough, USA) three times to visualize and quantify particle concentrations and sizes and thereby a potential area coverage. On the basis of this assessment, a specific volume of the sample, ranging between 0.2 and 100% of the calculated volume for full area coverage, was filtered onto an aluminum oxide filter (Ø = 25 mm; pore size, 0.2 μm; Anodisc, Whatman, Germany). This step is essential to avoid a particle overload of the filter, which would lead to a total absorption of subsequent measurement by FTIR radiation and error in the resulting particle counts. Despite this precaution, the membrane of eight snow samples became clogged during filtration such that it was stopped and the filtered volume was recorded.

After drying in a desiccator for 2 days, the filter was placed onto the calcium fluoride window of an FTIR microscope, and an overview image was taken (×40 magnification). The particles on the filter were analyzed by a Hyperion 3000 μFTIR microscope equipped with a focal plane array (FPA) detector with 64 × 64 detector elements and connected to a TENSOR 27 spectrometer (Bruker Optics GmbH, Germany). A visible objective (×4 magnification) and an infrared (IR) objective (×3.5 magnification) along with an IR range of 3600 to 1250 cm⁻¹ was used for measurements performed with the software OPUS 7.5 (Bruker Optics). The measurement area of 20 × 20 FPA (14.1 by 14.1 mm) fields comprised the entire sample-filter surface and produced measurements of 400 single fields. Measurements were performed by a 3.5× IR objective in a transmission mode with 32 scans per FPA field without binning at a resolution of 8 cm⁻¹, allowing the detection of particles down to 11 μm in ~4.5 hours.

Detection of fibers

Because of the design of the particle analysis pipeline during FTIR image analysis, fibers can currently not be identified reliably with
this automated method. Still, to enable comparison with previous studies, the fibers present on each filter were counted and photographed using a stereomicroscope (×8 magnification; Olympus SZX16). The total number of fibers per filter was extrapolated to 1 liter of melted snow. The total number of fibers found in the blanks (n = 83) was subtracted from the sample results. In addition, the length of fibers of suspected anthropogenic origin was measured using CellSens Micro Imaging tools (×8 to ×32 magnification; Olympus, Germany).

Data analysis
The FTIR imaging data were automatically processed (24). Briefly, each spectrum in the measurement file was analyzed via two library functions (25). The FTIR imaging data were automatically processed (25). Still, to enable comparison with previous studies, the fibers present on each filter were counted and photographed using a stereomicroscope (×8 magnification; Olympus SZX16). The total number of fibers per filter was extrapolated to 1 liter of melted snow. The total number of fibers found in the blanks (n = 83) was subtracted from the sample results. In addition, the length of fibers of suspected anthropogenic origin was measured using CellSens Micro Imaging tools (×8 to ×32 magnification; Olympus, Germany).

SUPPLEMENTARY MATERIALS
Supplementary material for this article is available at http://advances.sciencemag.org/cgi/content/full/5/8/eaax1157/DC1
Fig. S1. Length frequency (%) of microfibers found in Arctic and European snow.
Note S1. Calculation of propagation of uncertainty of results regarding possible contamination.
Table S1. Concentration of MPs, fibers polymer types, and other particles detected in snow samples.
Table S2. Concentration of MPs, fibers polymer types, and other particles detected in snow samples as well as size frequency distribution and metadata.
Table S3. Number, size, and color of fibers measured in Arctic and European snow samples.
Table S4. Concentration, composition, and size of MPs detected in blank samples.

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