Evaluation of Glass Wool Filters and Hollow-Fiber Ultrafiltration Concentration Methods for qPCR Detection of Human Adenoviruses and Polyomaviruses in River Water

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Abstract Pathogenic human viruses cause over half of gastroenteritis cases associated with recreational water use worldwide. They are difficult to concentrate from environmental waters due to low numbers and small sizes. Rapid enumeration of viruses by quantitative polymerase chain reaction (qPCR) has the potential to improve water quality analysis and risk assessment. However, capturing and recovering these viruses from environmental water remain formidable barriers to routine use. Here, we compared the recovery efficiencies of human adenoviruses (HAdVs) and human polyomaviruses (HPyVs) from 10-L river water samples seeded with raw human wastewater (100 and 10 mL) using hollow-fiber ultrafiltration (HFUF) and glass wool filter (GWF) methods. The mean recovery efficiencies of HAdVs in river water samples through HFUF were 36 and 86 % for 100 and 10 mL of seeded human wastewater, respectively. In contrast, the estimated mean recovery efficiencies of HAdVs in river water samples through GWF were 1.3 and 3 % for 100 and 10 mL seeded raw human wastewater, respectively. Similar trends were also observed for HPyVs. Recovery efficiencies of HFUF method were significantly higher ($P < 0.05$) than GWF for both HAdVs and HPyVs.

Our results clearly suggest that HFUF would be a preferred method for concentrating HAdVs and HPyVs from river water followed by subsequent detection and quantification with PCR/qPCR assays.

Keywords Microbial source tracking · Fecal indicator bacteria · Enteric viruses · Virus concentration · Human adenoviruses · Human polyomaviruses

1 Introduction

Pathogens have been found in environmental water sources as a result of fecal pollution from wastewater treatment plants (WWTPs), storm water drains, defective septic systems, and run-off from livestock and wildlife (Abdelzaher et al. 2010; Sidhu et al. 2012; Ahmed et al. 2013). Identification of the source(s) of fecal pollution provides the first step in initiating remediation efforts and minimizing human health risks.

This can be achieved by applying rapid microbial source tracking (MST) tools to identify and quantify host-specific genes or markers targeting bacteria, protozoa, and viruses (Harwood et al. 2014).

Development of numerous MST markers has been reported in the literature (Harwood et al. 2014). Among the enteric viral markers, human adenoviruses (HAdVs) and human polyomaviruses (HPyVs) have been most widely used to detect human wastewater pollution in environmental waters (Fong et al. 2005; Hundesa et al. 2006; McQuaig et al. 2009). HAdVs are responsible for a wide array of diseases such as gastroenteritis,
respiratory infections, eye infections, acute hemorrhagic cystitis, and meningoencephalitis (Videla et al. 1998; Echavarría 2008). On the other hand, HPyVs are unique to humans and generally produce asymptomatic viruria, especially in immunocompromised people (Polo et al. 2004). HPyVs are frequently excreted in urine from healthy individuals. Due to the high abundances of HAdVs and HPyVs in human feces and urine, they have received significant attention as MST markers (Fong et al. 2005; Hundesa et al. 2006; McQuaig et al. 2009; Ahmed et al. 2016).

Generally, polymerase chain reaction (PCR) and quantitative PCR (qPCR)-based assays are used to detect and quantify these viral markers in environmental samples (McQuaig et al. 2005; Staley et al. 2012; Rusiñol et al. 2009). Enteric viruses need to be concentrated from environmental water samples prior to PCR/qPCR analysis. The most commonly used concentration methods are hollow-fiber ultrafiltration (HFUF) (Rodriguez-Diaz et al. 2009), ultracentrifugation (Nordgren et al. 2009), adsorption-elution-based protocol with glass wool filter (GWF) (Lambertini et al. 2008), and positively and negatively charged membranes (Katayama et al. 2002; Bennett et al. 2010). The ability to recover maximum numbers of viruses from various water matrices, however, can be highly variable depending on the concentration methods used (Haramoto et al. 2006; Dubois et al. 2007).

Enteric viruses are relatively difficult to concentrate from environmental waters due to their low occurrence and small size (Maier et al. 2008). Therefore, recovery of viruses from environmental waters requires filtration on the scale up to 100 L of sample depending on the magnitude of fecal pollution. Among the most commonly used concentration methods, HFUF has been used widely to recover viruses from environmental waters with recovery rates ranging from 50 to 90 % (Morales-Morales et al. 2003; Hill et al. 2005; Hill et al. 2007; Polaczyk et al. 2008). Sodocalcic GWF also offers a promising alternative as an adsorptive material for virus concentration. GWF has been used to concentrate viruses from human wastewater (Gantzer et al. 1997) and environmental waters (Hot et al. 2003; Ehlers et al. 2005). Albinana-Gimenez et al. (2009a) compared GWF and ultrafiltration cartridge to recover known quantities of HAdV 2 and John Cunningham polyomavirus (JCPyV) in source water and drinking water using quantitative PCR. Both methods produced similar recovery efficiencies for HAdV 2 (GWF 4.2 %, ultrafiltration 5.1 %) but ultrafiltration had higher efficiencies (19 %) for JCPyV compared to GWF (4.4 %). Based on the results, the authors concluded that the GWF method produced acceptable and reproducible recovery efficiencies, whereas the ultrafiltration method yielded variable recovery efficiencies. It has to be noted that Albinana-Gimenez et al. (2009b) seeded cultured human adenoviruses type 2 (HAdV 2) and JCPyV obtained from plaque assays. In a real-world scenario, fecal pollution of environmental waters would occur via human wastewater/septic overflows. Little has been documented on the recovery efficiencies of HFUF and GWF methods for concentrating HAdVs and HPyVs markers from environmental water samples seeded with raw human wastewater.

The aim of this study was to compare the performance of HFUF and GWF concentration methods to recover HAdVs and HPyVs from river water samples. qPCR assays were used to measure the concentrations of these viruses in deoxyribonucleic acid (DNA) samples extracted from river water seeded with human wastewater.

2 Materials and Methods

2.1 Sample Preparation

A two-liter human wastewater sample was collected from the primary influent of a wastewater treatment plant (WWTP). The WWTP has a flow capacity of 54 megaliter day$^{-1}$ and treats human wastewater from approximately 250,000 people. The treatment process consists of primary treatment, secondary treatment (activated sludge), and disinfection with chlorine and ultraviolet (UV) prior to being discharged into the Brisbane River. A 100-L river water sample (clear color) was collected from the upstream in the Brisbane River in carboys at a depth 0.5 to 1 m. This site receives overflow of water from the Wivenhoe Reservoir after precipitation. The suspected sources of fecal pollution in this site include wildlife. Human wastewater and river water samples were stored at 4 °C for no more than 3 h before processing.

For each separate trial ($n = 3$), 100 and 10-mL volumes of human wastewater samples were seeded into two batches (9.9 and 9.99 L) of river water samples in triplicate (final volume of 10 L). The pH and turbidity of the river water sample were 8.0 ± 0.1 and 5.2 ± 0.3.
Nephelometric Turbidity Unit (NTU), respectively. The background concentrations of HAdVs and HPyVs in human wastewater and river water samples were also determined using qPCR assays (see below for details methodology). In brief, 100-mL raw human wastewater samples \((n = 3)\) were amended with HCl followed by passing through the 0.45-µm HA negatively charged 90-mm membranes (HAWP09000; Merck Millipore Ltd, Sydney, Australia). Two-milliliter DNA was extracted from the membranes using a PowerMax® Soil DNA Isolation Kit (MO BIO Laboratories, Inc., Carlsbad California, USA). River water \((n = 3)\) samples were concentrated using the HFUF method described below in details.

2.2 HAdVs and HPyVs Concentration and DNA Extraction

HAdVs and HPyVs were concentrated using a tangential flow HFUF method (Hill et al. 2005). General procedures are shown in Fig. 1. The method involves concentrating water samples using a Hemoflow HF80S dialysis filter (Fresenius Medical Care, Bad Homberg, Germany). Briefly, each 10-L water sample was pumped with a peristaltic pump in a closed loop with high-performance, platinum-cured L/S 36 silicone tubing (Masterflex; Cole-Parmer Instrument Co., Chicago, Illinois, USA). At the end of the concentration process, pressurized air was passed through the filter cartridge from the top to recover approximately 100 to 150-mL concentrated sample in the retentate container. To improve recovery, after each sample was processed through the HFUF, 500 mL of a surfactant solution (0.01 % Tween 80, [Sigma-Aldrich, St. Louis, Missouri] 0.01 % NaPP, and 0.001 % Antifoam A [Sigma-Aldrich, St. Louis, Missouri, USA]) was recirculated through the filter until the system started to draw air. This elution solution was collected and added to the retentate to achieve a final volume of approximately 250–300 mL and stored at 4 °C. A new filter cartridge was used for each sample. The pH of the concentrated sample was adjusted to 3.5 using 2.0 N HCl. The sample was then passed through 0.45-µm HA negatively charged 90-mm membranes (HAWP09000; Merck Millipore Ltd, Sydney, Australia) (McQuaig et al. 2009) attached to a glass membrane holder (Merck Millipore Ltd., Sydney, Australia). DNA was extracted from the membrane using a PowerMax® Soil DNA Isolation Kit (MO BIO Laboratories, Inc., Carlsbad California, USA) with slight modification (Gyawali et al. 2015).

![Fig. 1](https://example.com/figure1.png) Procedures for hollow-fiber ultrafiltration and glass wool filters methods for virus concentration from river water samples seeded with human wastewater.
Extracted DNA was eluted through the spin filter membranes by adding 2 mL Solution C6 and stored at \(-20^\circ C\) until processed.

The GWF method involves preparing glass wool as described elsewhere (Millen et al. 2012). General procedures are shown in Fig. 1. Washed glass wool was packed into cam lever couplings (Banjo, Crawfordsville, Indiana, USA) using a metal plunger. The cam lever coupling packed glass wool was flushed with 60 mL of sterile phosphate-buffered saline (pH = 6.8) using a catheter-tipped syringe. A water sample with a pH > 7.5 was adjusted to pH 6.5–7.0 by adding 2.0 N HCl. Each human wastewater-seeded water sample (10 L) was passed through the glass wool using a peristaltic pump. Viruses were eluated in the opposite direction to the original flow with 2 × 80 mL 3 % beef extract in 0.05 M glycine buffer with a pH of 9.5. The first eluent was allowed to soak the glass wool for 15 min before adding the second eluent, which was immediately pushed through the filter by air to obtain approximately 140 to 190-mL concentrated sample. The pH of eluate was adjusted to 7.0–7.5 using 2.0 N HCl. Concentrated sample (140 to 190-mL) was further concentrated using a Jumbosep™ (molecular weight cut-off = 100 kDa) Centrifugal Device (Pall Corporation, East Hills, NY, USA) to obtain a final volume of 5–6 mL. DNA (2 mL) was extracted directly from the 5–6-mL concentrated sample using a PowerMax® Soil DNA Isolation Kit (MO BIO Laboratories, Inc., Carlsbad California, USA) with slight modification as described earlier. To assess for cross-contamination, one negative process control (10 L of unseeded tap water) was processed for each method in parallel to river water samples seeded with human wastewater.

2.3 PCR Inhibition

A Sketa22 real-time PCR assay was performed to determine the level of PCR inhibition in DNA samples extracted from river water seeded with human wastewater (Haugland et al. 2005; Ahmed et al. 2015). River water DNA samples were seeded with a known amount (10 pg) of *Oncorhynchus keta* DNA (Sigma Chemical Co., St. Louis, MO). DNase- and RNase-free water samples were also seeded with 10 pg *O. keta* DNA. The threshold cycle ($C_T$) values for seeded *O. keta* DNA (10 pg) were determined in a PCR run for both river water DNA samples and DNase- and RNase-free water. The *O. keta* $C_T$ values obtained for DNase- and RNase-free water were compared to the $C_T$ values obtained for river water DNA samples to obtain information on the PCR inhibition level. A 2 $C_T$ delay was considered as having PCR inhibitors.

2.4 qPCR Standards and Assays

The HAdVs- and HPyVs-positive controls (DNA) were isolated from adenovirus strain 41 (ATCC VR-930) and raw human wastewater, respectively. The PCR-amplified products were purified using a QIAquick PCR Purification Kit (Qiagen, Valencia, California, USA) and cloned into a pGEM-T Easy Vector System II (Promega, Madison, Wisconsin, USA). Recombinant plasmids with corresponding inserts were purified using a Plasmid Mini Kit (Qiagen, Valencia, California, USA). Standards for qPCR assays of HAdVs and HPyVs were prepared from the plasmid DNA, ranging from $3 \times 10^5$ to $3 \times 10^1$ (for HAdVs) and $5 \times 10^5$ to $5 \times 10^0$ (for HPyVs) copies. The amplification efficiency ($E$) was determined by analysing the standards and was estimated from the slope of the curve as $E = 10^{-\text{slope}}$.

qPCR assays were performed using previously published assays using the Bio-Rad CFX96 thermal cycler (Bio-Rad Laboratories, Richmond, California, USA). The primer and probe sequences, concentrations, qPCR reaction volumes, and cycling parameters are shown in Table 1. Sketa22 real-time PCR amplifications were performed in 25-μL reaction mixtures containing 2 μL (10 pg) of *O. keta* DNA using iQ Supermixes (Bio-Rad Laboratories, Richmond, California, USA). HAdVs qPCR amplifications were performed in 20-μL reaction mixtures containing 3-μL DNA samples using SsoFast EvaGreen Supermix (Bio-Rad Laboratories, Richmond, California, USA), and HPyVs amplifications were performed in 50-μL reaction mixtures containing 5-μL DNA samples using TaqMan Universal PCR master mix (Applied Biosystems, Foster City, California, USA). All qPCR reactions were performed in triplicate. To minimize qPCR contamination, DNA extraction and qPCR set-up were performed in separate laboratories. A method blank was included for each batch of tap and river water samples. A reagent blank was also included during
DNA extraction. For each qPCR experiment, corresponding positive (standards) and negative controls (DNase- and RNase-free water) were included.

2.5 qPCR Assays Lower Limit of Quantification

The qPCR lower limit of quantification (LLOQ) was determined from the CT values obtained for each standard. To determine qPCR LLOQ, tenfold serial dilutions of standards (3 × 10^5 to 3 copies for HAdVs and 5 × 10^5 to 5 copies for HPyVs) were tested in triplicates. The minimum concentration of copies from the standard series detected in 3/3 qPCR reactions was considered qPCR LLOQ.

2.6 Recovery Efficiency and Statistical Analysis

The recovery efficiency of virus concentration method was calculated as follows:

\[
\text{Recovery efficiency } \% = \left( \frac{\text{concentration of copies recovered}}{\text{concentration copies seeded for each virus}} \right) \times 100
\]

A paired t test for equal means was conducted to determine the difference between HAdVs and HPyVs concentrations obtained through HFUF and GWF methods.

3 Results

3.1 Concentrations of HAdVs and HPyVs in Raw Human Wastewater and River Water Samples

The mean concentrations of HAdVs and HPyVs in raw human wastewater sample were 5.0 × 10^5 ± 5.0 × 10^4 copies and 3.2 × 10^5 ± 2.0 × 10^4 copies 100 mL^-1, respectively. None of the viruses were detected in 1 L of river water sample.

3.2 qPCR Standards and Lower Limit of Quantification

The standards had a linear range of quantification from 3 × 10^5 to 3 × 10^1 (for HAdVs) and 5 × 10^5 to 5 × 10^0 (for HPyVs) copies μL^-1 of DNA extract. The slope of the standards ranged from 3.306 to 3.422 (for HAdVs) and 3.239 to 3.377 (for HPyVs). The amplification efficiencies ranged from 96 to 101 % (for HAdVs) and 102 to 103 % (for HPyVs), and the correlation coefficient (r^2) ranged from 0.98 to 0.99 (for both HAdVs and HPyVs). Lower limit of quantification of qPCR assays was determined using the standards. The qPCR lower

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**Table 1** Target, primer/probe sequences and concentrations, and amplification conditions for endpoint PCR and qPCR assays used in this study

| PCR/qPCR assay | Target gene | Size of amplicons | Primers/probes sequence | Primers/probes concentrations | Cycling parameters | Reference |
|---------------|-------------|-------------------|-------------------------|-------------------------------|-------------------|----------|
| Sketa22       | ITS-2       | 77 bp             | F, GGT TTC CGC AGC TGG G R, CCG AGC CGT CCT GGT CTA P, FAM-AGT CGC AGG CGG CCA CCG T-TAMRA | 300 nM 300 nM 400 nM | 10 min at 95 °C, followed by 40 cycles of 15 s at 95 °C and 45 s at 63 °C | Haugland et al. 2005 |
| HAdVs         | Hexon       | 132 bp            | F, GCC ACG GTG GGG TTT CTA AAC AT R: GCC CCA GTG GTC TTA CAT GCA CAT C | 250 nM 250 nM | 10 min at 95 °C, followed by 40 cycles of 15 s at 95 °C, 20 s at 60 °C, and 20 s at 95 °C | Heim et al. 2003 |
| HPyVs         | Homologous T antigen | 176 bp | F, AGT CTT TAG GGT CTT CTA CCT TT R, GGT GCC AAC CTA TGG AAC AG P, FAM-AGT CGC AGG CGG CCA CCG T-MGBNFQ | 500 nM 500 nM 400 nM | 10 min at 95 °C, followed by 40 cycles of 15 s at 95 °C, 15 s at 55 °C, and 60 s at 60 °C | McQuaig et al. 2009 |
The limits of quantification were 30 and 5 copies for HAdVs and HPyVs, respectively.

### 3.3 PCR Inhibition Test

The mean $C_T$ value and standard deviation for the *O. keta*-seeded DNase- and RNase-free water were $28.0 \pm 0.3$. $C_T$ values for *O. keta*-seeded river water samples were comparable to DNase- and RNase-free water for both HFUF ($C_T = 28.4 \pm 0.3$) and GWF ($C_T = 28.6 \pm 0.5$) methods, suggesting the samples were potentially PCR inhibitors free.

### 3.4 Recovery Efficiencies of HAdVs and HPyVs

The mean recovered HAdV copy numbers in 10 L river water seeded with 100 mL ($1.6 \times 10^5 \pm 1.0 \times 10^4$ copies) and 10 mL ($2.0 \times 10^4 \pm 3.2 \times 10^3$ copies) human wastewater obtained through HFUF were much higher than those river water samples obtained through GWF ($C_T = 28.4 \pm 0.3$) and GWF ($C_T = 28.6 \pm 0.5$) methods, suggesting the samples were potentially PCR inhibitors free.

HFUF outperformed GWF in terms of recovering efficiencies of both HAdVs and HPyVs.

### 4 Discussion

HFUF method involved concentrating viruses from 10 L river water to obtain a manageable volume (250–300 mL) of sample. Consequently, the pH of the concentrated sample was adjusted to 3.5 (below the isoelectric point of the viruses). The concentrated samples were further passed through negatively charged HA membranes which adsorb the positively charged viruses present in the sample. HFUF has been used for concentrating HPyVs, human *Bacteroides* HF183, and *Methanobrevibacter smithii* from 10 L of environmental water samples seeded with 5 or 2.5 mL raw human wastewater (Leskinen et al. 2010). Leskinen et al. (2010) extracted DNA directly from the 47-mm membranes using a PowerSoil® DNA Isolation Kit. One limitation of using smaller diameter membranes is that they may get clogged depending on the turbidity of the water. As a result, multiple membranes may be required which may increase the sample processing time and may reduce recovery efficiency. In addition, DNA extraction using a PowerSoil® DNA Isolation Kit does not utilize all supernatant which may also influence the recovery efficiency in the downstream application. In view of these, we processed river water samples through 0.45-μm, 90-mm diameter negatively charged membranes. The larger diameter membranes provide larger net area (4.5 times more than 47-mm membranes) which allow to process relatively large volume of concentrated water samples (250–300 mL). For DNA extraction, we used PowerMax® Soil DNA Isolation Kit which can easily accommodate 90-mm diameter membrane. Unlike PowerSoil® DNA Isolation Kit, PowerMax® Soil Kit utilizes all supernatant and because of that better recovery is expected (Ahmed et al. 2015). A limitation of such approach can be the presence of potential PCR inhibitors on the membranes. However, the PowerMax® Soil Kit is equipped with inhibitor removal technology which potentially removes 100 % humic substances and other inhibitors as specified in the manual. This was supported by the Sketa22 real-time PCR assay which indicated the absence of PCR inhibitors in river water samples seeded with human wastewater obtained through both methods. It should be noted that 2-mL DNA sample was obtained using the PowerMax® Soil

![Fig. 2 HAdVs and HPyVs copies recovered from 10-L river water samples seeded with human wastewater](image)
Kit which is 10–20 times higher than PowerSoil® DNA Isolation Kit or other commercially available DNA extraction kits such as Qiagen Stool kit. This may reduce sensitivity of the qPCR assay. However, to increase qPCR sensitivity, 2-mL DNA can be reconcentrated further to a suitable volume (if required).

Between the two methods, HFUF had higher recovery than GWF for river water samples. These results were consistent for both viruses. The mean recovery efficiencies of 36 % (HAdVs) and 90 % (HPyVs) of HFUF obtained in this study can be considered quite efficient for simultaneous detection or quantification of these two viral markers in environmental waters. Our findings are in accordance with Morales-Morales et al. 2003; Hille et al. 2005; 2007; and Polaczyk et al. 2008 who also reported that HFUF can be effective for higher recovery (50–90 %) of viruses from various water matrices. In addition, HFUF is rapid and it does not require the preparation of extensive chemicals. The method also simultaneously retains bacteria, protozoa, and viruses in a single step which is an added advantage when analysis of multiple pathogens is required (Kfir et al. 1995; Oshima 2001; Morales-Morales et al. 2003).

GWF method used in this study was originally developed for detecting viruses and later it was used to detect agricultural zoonotic pathogens in large volume of drinking water, surface water, groundwater, and agricultural runoff (Vilagines et al. 1993; Millen et al. 2012). This approach has been successfully used to provide information of the recovery efficiencies of bacteria and viruses in 20 L of environmental water samples amended with variable amount of agricultural soil (Abd-Elmaksoud et al. 2014). The recovery efficiencies of bovine coronavirus, bovine rotavirus, and bovine viral diarrhea virus type 1 and type 2 using GWF have been reported to range from 13 to 26 % (at turbidity 0.5 NTU), 9 to 23 % (turbidity 215 NTU), and 14 to 24 % (turbidity 447 NTU), respectively. However, the standard deviations for each water type and virus varied significantly. The mean recovery efficiencies of HAdVs and HPyVs in this study for GWF ranged from 1.3 to 3.4 % for river water samples. Our findings are in accordance with Albinana-Gimenez et al. 2009b who also reported 4.2 % HAdVs and 4.4 % HPyVs recovery through GWF method. Francy et al. 2013 also reported 4.7 % recovery of HAdVs from lake water samples. Lower recovery efficiencies of other viruses such as H1N1 influenza and feline calcivirus F9 through the GWF method have been reported (Gassilloud et al. 2003; Deboosere et al. 2011). Caution should be exercised when comparing published studies on recovery efficiency of viral concentration methods as several factors such as adsorption of viruses to glass wool filters, glass wool filter type, seeding materials (raw human wastewater vs. intact plaques), environmental sample type (ground water vs. river water), sample volume (10 L vs. 50 L), and sensitivity of qPCR assays can influence recovery efficiencies (Bofill-Mas et al. 2006; Albinana-Gimenez et al. 2009a; Li et al. 2010).

The low recovery efficiencies of HAdVs and HPyVs could be due to the fact that viruses adsorbed on the glass wool filters were not effectively eluted. This assumption needs to be tested by directly extracting DNA from a segment of glass wool. The amount of glass wool packed into cam lever couplings may have affected the recovery efficiencies. For example, Vilagines et al. 1993 used 50-g glass wool which resulted in 75 % poliovirus 1 recovery compared to Menut et al. (1993) who used 5-g glass wool, which resulted in 25.5 % recovery. Glass wool filters are currently packed by hand and not commercially available in a column format. This makes QA/QC difficult as packing may vary from person to person, resulting in large variability in results (Cashdollar and Wymer 2013). In addition, viruses in the elution buffer also underwent an additional concentration step in a JumboSep before DNA extraction. Reconcentration of viruses is commonly used approach because the levels of viruses in environmental waters could be low. Reconcentration methods have some

### Table 2 Recovery efficiencies (mean ± standard deviation) of copy numbers of HAdVs and HPyVs from Brisbane River water samples seeded with raw human wastewater

| Methods | Volume of human wastewater seeded | HAdV % recovery | HPyV % recovery |
|---------|-----------------------------------|----------------|-----------------|
| HFUF    | 100 mL                            | 36 ± 2.7       | 90 ± 4          |
|         | 10 mL                             | 86 ± 15        | 54 ± 13         |
| GWF     | 100 mL                            | 1.3 ± 0.4      | 3.4 ± 0.5       |
|         | 10 mL                             | 3.0 ± 1.0      | 0.01            |
disadvantages, such as these methods do not produce consistent recovery efficiency for different viruses (Lewis and Metcalf 1988). In a previous study, Centriprep Filter Concentrators provided high and stable recovery yields (74 %) of polioviruses (Haramoto et al. 2004). Another study reported the 35 % recovery of adenovirus 41 through Centricron filters (Wu et al. 2011).

To determine the recovery efficiency, we seeded raw human wastewater compared to the other studies that seeded cultured viruses obtained from plaque assays (Hill et al. 2005; Albinana-Gimenez et al. 2009b; Millen et al. 2012; Francy et al. 2013; Abd-Elmaksoud et al. 2014). In real-world scenario, fecal pollution of environmental waters occurs via sewer/septic overflows and surface run off containing fecal matters from various animals. Therefore, it is deemed necessary to obtain recovery efficiency of HAdVs and HPyVs by seeding raw human wastewater that potentially contains naked (genetic materials from defective virions) and intact viral genomes than other studies which seeded intact viruses. From the MST context, there is no need to elute “intact” viruses since the objective is to determine whether the sample contains the signature of human fecal pollution.

5 Conclusions

Our data suggest that HFUF method provides better recovery for HAdVs and HPyVs compared to GWF method used in this study. Therefore, we recommended that HFUF method should be used to concentrate water samples for MST field studies. The advantage of HFUF method is that these filters are readily available for use in dialysis treatment of patients, and sample-processing time is relatively shorter than the GWF method. Although, low turbid river water samples were tested in this study, the results may be applicable to stormwater, drinking water, and recreational water.

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References

Abd-Elmaksoud, S., Spencer, S. K., Gerba, C. P., Tamimi, A. H., Jokela, W. E., & Borchardt, M. A. (2014). Simultaneous concentration of bovine viruses and agricultural zoonotic bacterial from water using sodocalcic glass wool. Food Environ Virol, 6(4), 104–109.

Abdelzaher, A. M., Wright, M. E., Ortega, C., Solo-Gabriele, H. M., Miller, G., Elmir, S., Newman, X., Shih, P., Bonilla, J. A., Bonilla, T. D., Palmer, C. J., Scott, T., Lukasik, J., Harwood, V. J., McQuag, S., Sinigalliano, C., Gidley, M., Plano, L. R. W., Zhu, X., Wang, J. D., & Fleming, L. E. (2010). Presence of pathogens and indicator microbes at a non-point source subtropical recreational marine beach. Applied and Environmental Microbiology, 76(3), 724–732.

Ahmed, W., Sritharan, T., Palmer, A., Sidhu, J. P. S., & Toze, S. (2013). Evaluation of bovine feces-associated microbial source tracking markers and their correlations with fecal indicators and zoonotic pathogens in a Brisbane, Australia, reservoir. Applied and Environmental Microbiology, 79(8), 2682–2691.

Ahmed, W., Harwood, V. J., Gyawali, P., Sidhu, J. P. S., & Toze, S. (2015). Comparison of concentration methods for quantitative detection of sewage-associated viral markers in environmental waters. Applied and Environmental Microbiology, 81(6), 2042–2049.

Ahmed, W., Sidhu, J. P. S., Smith, K., Beale, D., Gyawali, P., & Toze, S. (2016). Distributions of fecal markers in wastewater from varying climatic zones for human fecal pollution tracking in Australian surface waters. Applied and Environmental Microbiology, 82(4), 1316–1323.

Albinana-Gimenez, N., Miagostovich, M. P., Calgua, B., Huguet, J. M., Matia, L., & Girones, R. (2009a). Analysis of adenoviruses and polyomaviruses quantified by qPCR as indicators of water quality in source and drinking-water treatment plants. Water Research, 43(7), 2011–2019.

Albinana-Gimenez, N., Clemente-Casares, P., Calgua, B., Courtois, S., & Girones, R. (2009b). Comparison of methods for concentrating human adenoviruses, polyomavirus JC, and noroviruses in source waters and drinking waters using quantitative PCR. Journal of Virological Methods, 158(1-2), 104–109.

Bennett, H. B., O’Dell, H. D., Norton, G., Shin, G., Hsu, F. C., & Meschke, J. S. (2010). Evaluation of a novel electropositive filter for the concentration of viruses from diverse water matrices. Water Science and Technology, 61(2), 317–322.

Bofill-Mas, S., Albinana-Gimenez, N., Clemente-Casares, P., Hundesa, A., Rodriguez, M. J., Allard, A., Calvo, M., & Girones, R. (2006). Quantification and stability of human adenoviruses and polyomavirus JC/CyPyV in wastewater matrices. Applied and Environmental Microbiology, 72(12), 7894–7896.

Cashdollar, J. L., & Wymer, L. (2013). Methods for primary concentration of viruses from water samples: a review and meta-analysis of recent studies. Journal of Applied Microbiology, 115(1), 1–11.

Deboosere, N., Horn, S. V., Pinon, A., Gachet, J., Coldefy, C., Buchy, P., & Vialette, M. (2011). Development and validation of a concentration method for the detection of influenza A viruses from large volumes of surface water. Applied and Environmental Microbiology, 77(11), 3802–3808.

Dubois, E., Hennecart, C., Merle, G., Burger, C., Hmila, N., Ruelle, S., Perelle, S., & Ferré, V. (2007). Detection and quantification by real-time RT-PCR of hepatitis A virus from inoculated tap waters, salad vegetables, and soft fruits.
characterization of the method performances. International Journal of Food Microbiology, 117(2), 141–149.

Echavarría, M. (2008). Adenoviruses in immunocompromised hosts. Clinical Microbiology Reviews, 21(4), 704–715.

Ehlers, M. M., Grabow, W. O., & Pavlov, D. N. (2005). Detection of enteroviruses in untreated and treated drinking water supplies in South Africa. Water Research, 39(11), 2253–2258.

Fong, T. T., Griffin, D. W., & Lipp, E. K. (2005). Molecular assays for targeting human and bovine enteric viruses in coastal waters and their application for library-independent source tracking. Applied and Environmental Microbiology, 71(4), 2070–2078.

Francy, D. S., Stelzer, E. A., Brady, A. M. G., Huitle, C., Bushon, R. N., Ip, H. S., Water, M. W., Villeegas, E. N., Gallardo, V., & Alan Lindquist, H. D. (2013). Comparison of filters for concentrating microbial indicators and pathogen in lake water samples. Applied and Environmental Microbiology, 79(4), 1342–1352.

Gantzer, C., Senouci, S., Maul, A., Levi, Y., & Schwartzbrod, L. (1993). Poliovirus recovery in untreated and treated drinking water supplies in South Africa. Water Research, 39(11), 2253–2258.

Gassilloud, B., Duval, M., Schwartzbrod, L., & Gantzer, C. (1993). Recovery of feline calicivirus infectious particles and genome from water: comparison of two concentration technique. Water Science and Technology, 47(3), 97–101.

Gyawali, P., Ahmed, W., Jagals, P., Siddhu, J. P. S., & Toze, S. (2015). Comparison of concentrations methods for rapid detection of ova in wastewater matrices using quantitative PCR. Environmental Parasitology, 159, 160–167.

Haramoto, E., Katayama, H., & Ohgaki, S. (2004). Detection of noroviruses in tap water in Japan by means of a new method for concentrating enteric viruses in large volumes of freshwater. Applied and Environmental Microbiology, 70(4), 2154–2160.

Haramoto, E., Katayama, H., Oguma, K., Yamashita, H., Tajima, A., Nakajima, H., & Ohgaki, S. (2006). Seasonal profiles of human noroviruses and indicator bacteria in a wastewater treatment plant in Tokyo, Japan. Water Science and Technology, 54(11), 301–308.

Harwood, V. J., Staley, C., Badgely, B. D., Borges, K., & Korajkic, A. (2014). Microbial source tracking markers for detection of fecal contamination in environmental waters: relationships between pathogens and human health outcomes. FEMS Microbiology Review, 38(1), 1–40.

Haugland, R. A., Siefring, S. C., Wymer, L. J., Brenner, K. P., & Dufour, A. P. (2005). Comparison of Enterococcus measurements in freshwater at two recreational beaches by quantitative polymerase chain reaction and membrane filter culture analysis. Water Research, 39(4), 559–568.

Heim, A., Ebnert, C., Harste, G., & Pring-Åkerblom, P. (2003). Rapid and quantitative detection of human adenovirus DNA by real-time PCR. Journal of Medical Virology, 70(2), 228–239.

Hill, V. R., Polaczek, A. L., Hahn, D., Jothikumar, N., Cromeans, T. L., Roberts, J. M., & Amburgey, J. E. (2005). Development of a rapid method for simultaneous recovery of diverse microbes in drinking water by ultrafiltration with sodium polyphosphate and surfactants. Applied and Environmental Microbiology, 71(11), 6878–6884.

Hill, V. R., Kahler, A. M., Jothikumar, N., Johnson, T. B., Hahn, D., & Cromeans, T. L. (2007). Multistate evaluation of an ultrafiltration-based procedure for simultaneous recovery of enteric microbes in 100-liter tap water samples. Applied and Environmental Microbiology, 73(13), 4218–4225.

Hot, D., Leganey, O., Jacques, J., Gantzer, C., Caudrelier, Y., Guyard, K., Lange, M., & Andreoletti, L. (2003). Detection of somatic phages, infectious enteroviruses and enterovirus genomes as indicators of human enteric viral pollution in surface water. Water Research, 37(19), 4703–4710.

Hundesa, A., Maluque de Motes, C., Boiffil-Mas, S., Albinana-Gimenez, N., & Girones, R. (2006). Identification of human and animal adenoviruses and polyomaviruses for determination of sources of fecal contamination in the environment. Applied and Environmental Microbiology, 72(12), 7886–7893.

Katayama, H., Shimasaki, A., & Ohgaki, S. (2002). Development of a virus concentration methods and its application to detection of enterovirus and Norwalk virus from coastal seawater. Applied and Environmental Microbiology, 68(3), 1033–1039.

Kflr, R., Hilmer, C., Freez, M. D., & Bateman, B. (1995). Studies evaluating the implications of utilizing the same concentration techniques for the detection of protozoan parasites and viruses in water. Water Science and Technology, 31(5–6), 417–423.

Lambertini, E., Spencer, S. K., Bertz, P. D., Loge, F. J., Kieke, B. A., & Borchardt, M. A. (2008). Concentration of enteroviruses, adenoviruses, and noroviruses from drinking water by use of glass wool filters. Applied and Environmental Microbiology, 74(10), 2990–2996.

Leskienen, S. D., Brownell, M., Lim, D. V., & Harwood, V. J. (2010). Hollow-fiber ultrafiltration and PCR detection of human-associated genetic markers from various types of surface waters in Florida. Applied and Environmental Microbiology, 76(12), 4116–4117.

Lewis, D., & Metcalf, T. G. (1988). Removal of viruses in sewage treatment: assessment of feasibility. Microbiological Sciences, 5(9), 260–264.

Li, D., Shi, H.-C., & Jiang, S. C. (2010). Concentration of viruses from environmental waters using nanoalumina fiber filters. Journal of Microbiological Methods, 81(1), 33–38.

Maier, R. M., Pepper, I. L., Gerba, C. P. (2008). Environmental microbiology. San Diego: Academic Press.

McQuaig, S. M., Scott, T. M., Lukasik, J. O., Paul, J. H., & Harwood, V. J. (2009). Quantification of human polyomaviruses JC virus and BK virus by TaqMan quantitative PCR and comparison to other water quality indicators in water and fecal samples. Applied and Environmental Microbiology, 75(11), 3379–3388.

McQuaig, S., Griffith, J., & Harwood, V. J. (2012). Association of fecal indicator bacteria with human viruses and microbial source tracking markers at coastal beaches impacted by non-point source pollution. Applied and Environmental Microbiology, 78(18), 6423–6432.

Menut, C., Beril, C., & Schwartzbrod, L. (1993). Poliovirus recovery from tap water after concentration over glass powder and glass wool. Water Science and Technology, 27(3–4), 291–294.

Millen, H. T., Gonnering, J. C., Berg, R. K., Soencer, S. K., Jokela, W. E., Pearce, J. M., Borchardt, J. S., & Borchardt, M. A. (2012). Glass wool filters for concentrating waterborne viruses and agricultural zoonotic pathogens. Journal of Visualized Experiments, 61, e3930.
Optimization of a reusable hollow fiber ultrafiltration for simultaneous concentration of enteric bacteria, protozoa, and viruses from water. *Applied and Environmental Microbiology*, 69(7), 4098–4102.

Nordgren, J., Matussek, A., Mattsson, A., Svensson, L., & Per-Eric Lindgren, P. (2009). Prevalence of norovirus and factors influencing virus concentrations during one year in a full-scale wastewater treatment plant. *Water Research*, 43(4), 1117–1125.

Oshima, K. H. (2001). Efficient and predictable recovery of viruses and *Cryptosporidium parvum* oocysts from water by ultrafiltration systems (Technical Completion Report 01-4-23949). New Mexico State University, Las Cruces: New Mexico Water Research Resources Institute.

Polaczyk, A. L., Narayanan, J., Cromeans, T. L., Hahn, D., Roberts, J. M., Amburgey, J. E., & Hill, V. R. (2008). Ultrafiltration-based techniques for rapid and simultaneous concentration of multiple microbe classes from 100-L tap water. *Journal of Microbiological Methods*, 73(2), 92–99.

Polo, C., Perez, J. L., Mielnichuck, A., Fedele, C. G., Niuboad, J., & Tenorio, A. (2004). Prevalence and patterns of polyoma-virus urinary excretion in immunocompetent adults and children. *Clinical Microbiology and Infection*, 10(7), 640–644.

Rodriguez-Diaz, J., Quernles, L., Caraballo, L., Vizzi, E., Liprandi, F., Takiff, H., & Betancourt, W. Q. (2009). Detection and characterization of waterborne gastroenteritis viruses in urban sewage and sewage-polluted river waters in Caracas, Venezuela. *Applied and Environmental Microbiology*, 75(2), 387–394.

Rusinol, M., Fernandez-Cassi, X., Hundesa, A., Vieira, C., Kern, A., Eriksson, I., Ziros, P., Kay, D., Miagostovich, M., Vargha, M., Allard, A., Vantarakis, A., Wynn-Jones, P., Bofill-Mas, S., & Girone, R. (2014). Application of human and animal viral microbial source tracking tools in fresh and marine waters from five different geographical areas. *Water Research*, 59(1), 119–129.

Sidhu, J. P. S., Hodgers, L., Ahmed, W., Chong, M. N., & Toze, S. (2012). Prevalence of human pathogens and indicators in stormwater runoff in Brisbane, Australia. *Water Research*, 46(20), 6652–6660.

Staley, C., Gordon, K. V., Schoen, M. E., & Harwood, V. J. (2012). Performance of two quantitative PCR methods for microbial source tracking of human sewage and implications for microbial risk assessment in recreational waters. *Applied and Environmental Microbiology*, 78(20), 7317–7326.

Videla, C., Carballal, G., Misirlian, A., & Aguilar, M. (1998). Acute lower respiratory infections due to respiratory syncytial virus and adenovirus among hospitalized children from Argentina. *Clinical and Diagnostic Virology*, 10(1), 17–23.

Vilagines, P., Sarrette, B., Husson, G., & Vilagines, R. (1993). Glass wool for virus concentration at ambient water pH level. *Water Science and Technology*, 27(3-4), 299–306.

Wu, J., Rodriguez, R. A., Stewart, J. R., & Sobsey, M. D. (2011). A simple and novel method for recovering adenovirus 41 in small volumes of source water. *Journal of Applied Microbiology*, 110(5), 1332–1340.