INTRODUCTION

The pumpkinseed, *Lepomis gibbosus* (L.), is a telling example of successful fish introduction into the majority of European inland bodies of water (Crivelli and Mestre 1988, Holčik 1991, Welcomme 1991, Tomeček et al. 2007). The native environment of this species is being affected by the warm climate of eastern North America (Scott and Crossman 1973). It is thought that pumpkinseed was first imported around 1880 from Canada (Garcia-Berthou et al. 2005). Initially, *L. gibbosus* was released, as an ornamental fish, to garden ponds and aquaria (Copp et al. 2005a).

Later pumpkinseed was spread by re-stocking material of common carp (Tandon 1976). Nowadays, this species is widely dispersed in Central and Western Europe (Welcomme 1991), as well as in Iberian and Balkan Peninsulas (Elvira and Almodóvar 2001, Economidis et al. 2000). In northern areas, pumpkinseeds were found to be abundant only in artificial- or heavily disturbed water bodies (van Kleef et al. 2008).

The distribution, growth, and feeding behaviour of pumpkinseeds are subject to profound analyses, mainly with aim to assess the potential impact of this alien species on the native fauna.
species on native fish assemblages and the aquatic ecosystem as a whole (Crivelli and Mestre 1988, Fox and Crivelli 2001, Copp et al. 2004, Klaar et al. 2004, Villeneuve et al. 2005, Copp and Fox 2007, Nikolova et al. 2009). It was found that in most of the European ecosystems where pumpkinseeds have been introduced they influenced directly or indirectly native species through competition for food and predation on fish eggs and endemic molluscs (García-Berthou and Moreno-Amich 2000, Copp et al. 2005b), zooplankton (Brabrand and Saltveit 1989), and as well as macroinvertebrates (van Kleef et al. 2008, Nikolova et al. 2009).

In Bulgaria, pumpkinseed has been known since the mid-1920s when it was discovered in marshes along the Danube River (Drensky 1923). It is thought that the River Danube played a key role in pumpkinseed distribution, not only in Bulgaria, but also throughout Europe (Marinov 1966). Nearly 40 years later, several new populations of pumpkinseeds were reported in Bulgarian waters—mainly in tributaries and floodplain areas along the Danube River (Bulgurkov 1958, Karapetkova 1972, Karapetkova and Undzian 1988) and in some lowland reservoirs (Karapetkova 1972). Pumpkinseeds were not found in the river mouths of the Black Sea, except in the River Kamchja (Vassilev 1999). The current information about distribution and abundance of pumpkinseed in Bulgarian water bodies is scarce (Uzunova et al. 2008, Nikolova et al. 2009).

One of the key factors responsible for successful colonization of non-native habitats by pumpkinseeds are the flexible habitat requirements of this species (Klaar et al. 2004, Fox et al. 2007, Cucherousset et al. 2009). A high degree of variation in pumpkinseed’s population parameters (growth rate, reproduction, abundance, size, and age structure) can be a side effect of macro- and microhabitat characteristics and biotic conditions as well (food availability, predation, and food competition). It is considered that humans are not only the main vector in the introduction and dispersal of pumpkinseed (Copp et al. 2005a, García-Berthou et al. 2005, Tomeček et al. 2007), but even facilitate invasion process by alteration of water body habitats and water quality (van Kleef et al. 2008). However, it has remained difficult to isolate mechanisms influencing pumpkinseed’s population parameters. From a practical point of view, it would be very helpful to find the combination of those environmental factors responsible for success in the growing range and increased abundance of pumpkinseed and, on this basis, to develop and implement programs directed to promote the control of expansion of the invaders (Ricciardi and Rasmussen 1998, Kolar and Lodge 2001, Gozlan and Newton 2009, Panov et al. 2009, Gozlan et al. 2010).

The aim of this study was to improve our understanding of how biotic- and abiotic parameters as well as anthropogenic factors of investigated bodies of water interact with basic population parameters of non-native pumpkinseed. Specific research objectives of the study include: document the distribution, relative abundance, size, and age composition of L. gibbosus in the littoral zones of different type of water bodies in Bulgaria; evaluate relations between relative pumpkinseed abundance, age, size and environmental variables characterised macro and microhabitats. We selected 13 bodies of water—differing in their origin, use, water quality, size, depth, and the microhabitat characteristics of their littoral zones—for the investigation of the above-mentioned parameters and relations.

**MATERIAL AND METHODS**

**Study design.** Thirteen artificial, semi-artificial, and natural bodies of water were surveyed in 2010 in order to assess the pumpkinseed abundance, size, and age structure. According to their specific management aims of use, water bodies were divided into five main categories: 1 = active sand-pit lakes (n = 2), 2 = former sand-pit lakes (n = 4), 3 = reservoirs for cage aquaculture (n = 2), 4 = reservoirs for water supply (industrial and irrigation) (n = 3), 5 = waste waters deposition (adaptive ponds of refineries) (n = 2). All water bodies are used for recreational fishing. Macrohabitat parameters (maximum water depth, surface area, altitude) of the investigated bodies of water are shown in Table 1. Thirty-one survey sites located along the shoreline of the water bodies were chosen to represent a range of littoral habitats with variable degree of structural complexity. The selected sites were representative for the particular water body. Sampling was conducted during the period between the 10 May and 30 June 2010 that coincided with pumpkinseed breeding season.

**Sampling protocol.** Each study site was sampled with minnow trap-nets to estimate the relative abundance of pumpkinseed and other small-bodied fish inhabiting the littoral zone. Minnow traps were chosen because they could be set in almost any lake or habitat type (MacRae and Jackson 2006). We used commercially available Gee minnow traps with an opening diameter of 6–7 cm and a mesh size of 0.45 cm. Traps were baited with a dry trout feed and were set for 1 h during the sunny days between 0900 and 1600 h. The trap-nets were fixed to the shore with 5 m and 2.5 m cords in chessboard pattern. They were set along the coast, 2.5 m from each other. The number of traps used depended on the site length and varied from 15 to 20 per site. Collected fish were identified in situ (Kottelat and Freyhoff 2007). Pumpkinseeds were euthanized using clove oil, preserved in 10% formalin, then rinsed in water before being stored in 70% ethanol until analysed. Native fishes were released at the site after counting.

**Habitat assessment.** The following biotic-, and physical habitat variables at two transects parallel to the shore (between 1 and 5 m off the coast) were measured or scored: water depth, bottom substrate type, bank slope, percent submerged aquatic vegetation, type of coastal vegetation, electro-conductivity, oxygen saturation, dissolved oxygen, pH, and water temperature. At each study site, the physical and chemical parameters were measured 30 min after removal of the traps. Mean water depth and
substrate composition (in % of section area) were visually estimated at 1 m intervals at each transect and categorised as follows: silt (<0.06 cm), sand (0.06–0.2 cm), gravel (0.2–4.0 cm), boulders (>25 cm). Aquatic vegetation was scored 1, 2, 3, or 4 (1 = missing, 2 = sparse, 3 = intermediate, or 4 = dense). Coastal vegetation was described as: 1 = reed-belt, 2 = mixed broad-leaved, or 3 = grass. Bank slope received score 1 or 2 (1 = steep, 2 = gently inclined). Presence or absence of piscivore fish (= predators) was described in two categories: 1 = present and 2 = absent.

Electro-conductivity (µS · cm–1); oxygen saturation (ppm), dissolved oxygen (mg·L–1); pH and water temperature (°C) were measured in situ using a WTW Oxi 330i, WTW pH 330i and HI 98311.

**Fish analyses.** A total of 1950 pumpkinseed specimens were collected and their total length (TL) was measured to the nearest 0.01 cm. Fish age was determined from scales removed from below the anterior part of the dorsal fin. Scale measurements were made on the oral radius. The ageing study was conducted using a microfilm reader at magnification of 17.5 ×. Two investigators independently determined the age of fish. Age was confirmed if the percentage of disagreement was below 10 percent. For scales with higher discrepancies, additional measurements were made until agreement was reached.

**Data analyses.** Catch per unit effort (CPUE) was used as a measure of fish abundance. The CPUE was calculated as the number of fish caught per trap per hour [FTH]. Analysis of variance (one-way ANOVA, P < 0.05) was applied to test the effect of environmental characteristics on the pumpkinseed’s biological parameters and—in case of significance—Tukey post hoc tests were carried out to test for differences in mean values between the habitat types. Prior to analysis, all variables including fish abundance, habitat, and water quality measurements were evaluated for normality (normal probability plot or Chi-squared analysis). Those that significantly departed from a reasonable normal distribution were log (x + 1) (continuous variables) or arcsine square root (x · 100–1) transformed (percentages). Relations between pumpkinseeds abundance, size, age, and abundance of other fish species were tested using simple linear regression, following log (x + 1) transformations of the raw data. For statistical analysis, the programs Statistica 7.0 (StatSoft© Inc.) and XLStat (Addinsoft) were used.

**RESULTS**

A total of 11 fish species belonging to 4 families were caught in littoral zones of the investigated water bodies. *L. gibbosus* has been recorded at 26 sites, while 5 sites, located at the natural Lake Varnensko and the active sand-pit Lake Chepinci were devoid of pumpkinseeds. Actually, at two sites (CHP2 and VE), situated in the mentioned above water bodies we did not succeed in catching any fish. As a whole, species diversity of all investigated littoral zones was relatively low and varied between 1 and 4 species (Fig. 1). *L. gibbosus* dominated at 24 sites and constituted 86.7% of all fish caught. Pumpkinseed abundance and total abundance varied broadly from site to site (from 0 to 25.4 and from 0 to 37.1 FTH, respectively) (Fig. 2). The mean pumpkinseed abundance was 6.35 ± 7.2 FTH (mean ± standard deviation). The highest total abundance (37.4 FTH) was recorded at site located in Sopot Reservoir, followed by the former sand–pit lakes Svetovrachane and Negovan (Fig. 2). The lowest total abundance was observed at sites situated

**Table 1**

Principal characteristics of the Bulgarian bodies of water investigated for the abundance of pumpkinseed, *Lepomis gibbosus* L.

| Name of water body | Investigated sites (codes) | Geographical coordinates | Type of water body and use | Surface area [km²] | Maximum depth [m] | Altitude [m] |
|--------------------|---------------------------|--------------------------|---------------------------|-------------------|------------------|--------------|
| Dolni Bogrov       | DB1–DB4                   | 42°42′N, 23°28′E         | Former sand-pit lake      | 0.37              | 4.5              | 550          |
| Negovan            | NEGI–NEG2                 | 42°46′N, 23°23′E         | Former sand-pit lake      | 0.1               | 2.8              | 550          |
| Svetovrachene      | SV1–SV4                   | 42°46′N, 23°23′E         | Former sand-pit lake      | 0.04              | 2.2              | 550          |
| Chepintzi          | CHP1–CHP4                 | 42°44′N, 23°25′E         | Active sand-pit lake      | 0.24              | 19               | 560          |
| Chelopechene       | CHL1–CHL2                 | 42°43′N, 23°27′E         | Active sand-pit lake      | 0.82              | 15               | 550          |
| Kurdjali           | KR1–KR2                   | 41°38′N, 25°18′E         | Reservoir; cage aquaculture | 16.07             | 65               | 340          |
| Pleven (“Plama” refinery) | PL1–PL2             | 43°24′N, 24°30′E         | Waste water discharge     | 0.05              | 3                | 70           |
| Jasen              | JS1–JS2                   | 43°24′N, 24°30′E         | Waste water discharge     | 0.03              | 2.8              | 69           |
| Dospat             | DS1–DS3                   | 41°38′N, 24°09′E         | Reservoir; cage aquaculture | 22                | 74               | 1220         |
| Ognjanovo          | OG1–OG2                   | 42°36′N, 23°44′E         | Reservoir; water supply   | 4.16              | 47               | 630          |
| Marchaevo          | MR1–MR2                   | 42°35′N, 23°09′E         | Reservoir; water supply   | 0.01              | 3.2              | 770          |
| Sopot              | SOP                       | 42°59′N, 24°28′E         | Reservoir; water supply   | 14                | 28               | 400          |
| Varnensko          | VE                        | 43°11′N, 27°46′E         | Sea-coast lake; waste water discharge | 19                | 19               | 0            |

N = number of sites investigated.
Pumpkinseeds were the most numerous in the littoral zones of the former sand-pit lakes Negovan and Svetovrachane, where pumpkinseed's CPUE achieved values of 20 FTH. The sites with the lowest abundance were situated in the Kardjali Reservoir, active sand-pit lakes (Chelopechene and Chepintzi) and micro-reservoir Marchaevo. Two piscivore fish—European perch, *Perca fluviatilis* L., and pike perch, *Sander lucioperca* (L.)—were caught at 11 of the 31 study sites. Total piscivore abundance range from 0.1 to 24.8 FTH. In total, perch made up 97.1% of the total piscivores caught while pike perch made up only 2.9%. Bitterling, *Rhodeus sericeus* (Pallas, 1776), dominated at only one site (CHL1), but with very low relative abundance (1.0 FTH). One another exotic fish species, stone moroko, *Pseudorasbora parva* (Temmincket Schlegel, 1842), was found at five sites. The least frequently found (<0.1%) species that composed the littoral ichthyofauna in the investigated sites were: rudd, *Scardinius erythrophthalmus* (L.); common carp, *Cyprinus carpio* L.; Prussian carp, *Carassius gibelio* (Bloch, 1782); tench, *Tinca tinca* (L.); bleak, *Alburnus alburnus* (L.); and weatherfish, *Misgurnus fossilis* (L.).

The age of collected pumpkinseeds varied significantly between sites and water bodies from 0+ to 4+ (Fig. 3). The most numerous were fish at age 1+ and 2+. The highest mean pumpkinseed’s age was recorded at sites in lake for waste-water deposition (JS1 and JS2). In Sopot Reservoir, all captured pumpkinseeds were at age 1+. The mean length (TL) of pumpkinseeds ranged from 39 through 82 mm. Larger pumpkinseeds were found at sites located in the big reservoirs: Kardjali, Dospat, and Ognjanovo. The biggest specimen (120 mm) was caught at the site located in Dospat Reservoir (DOS3) and the smallest in Sopot Reservoir (Fig. 4).

Analysis of the effect of different macro- and micro-habitat environmental parameters on pumpkinseed’s abundance showed following relations (Tables 2 and 3). The majority of macrohabitat parameters did not affect any of population characteristics (Table 2). Only the type of water body had a significant effect on pumpkinseed abundance.
abundance and HDStestrevealedthatthelittoralzonesin
theactiveandsand-pitlakes,werecharacterizedbythelowest
number of pumpkinseeds than the other types of water
bodies (Tukey’s HSD, to all four types $P < 0.05$) (Fig. 5).
Microhabitat parameters such as silt and sand bottom sub-
strate also significantly affected pumpkinseeds abundance,
di andage, while gravel significantly affected
body length but not abundance and age (Table 3). Mean
pumpkinseed age was significantly different only for the
pair comparisons of former and active sand pit lakes
(Tukey HSD, $P < 0.001$) (Fig. 6). Pumpkinseeds with the
lowest mean lengths were found in the active sand-pit lakes
(Tukey HSD, to 2, 3, and 4 water body type $P < 0.05$) (Fig. 7).
Aquatic vegetation also had a significant impact on the
abundance and in sites with dense vegetation pumpkin-
seeds were more abundant than the sites with sparse vege-
tation (Tukey HSD, $P = 0.008$) (Fig. 8). The presence of
predator fishes was a significant factor influencing pump-
kinseed’s mean length (ANOVA, $F = 8.09, P = 0.008$). At
the sites where piscivore fish occurred, the mean length and
mean age of pumpkinseed were lower (ANOVA, $F = 8.09,
$ $P = 0.008; F = 10.4, P = 0.003$).

There was a significant relation, best described by lin-
ear regression, between pumpkinseed abundance and
species richness (Fig. 9). The regression line shows that
the abundance of pumpkinseeds increasing with an
increase of the number of fish species in the littoral zone
($r^2 = 0.4612, P = 0.0003$). There were no significant cor-
relations between total piscivore or individual piscivore
species abundance and pumpkinseed abundance.

**DISCUSSION**

The observed wide distribution of pumpkinseed in
Bulgarian water bodies did not differ significantly from
the situation in the majority of European countries, where
pumpkinseed can be found in a diverse range of water
body types including natural and semi-natural lakes, small
wetlands, reservoirs, marshes, fish ponds, and rivers
(Klaar et al. 2004, Oliva-Paterna et al. 2005, Tomeček et
al. 2007). And while the existence of pumpkinseeds is not
surprising for small water bodies located in a proximity to
urban areas, it is unusual for high mountain reservoirs
with low annual temperatures and deep waters, such as
Dospat Reservoir. The extremely high abundance of the
species in those waters can possibly be attributed to the
operation of intensive cage aquaculture facilities during
recent years, which steadily leads to high levels of
eutrophication of the basic oligotrophic artificial lakes.
Pumpkinseed was introduced in these reservoirs as a bait fish or with restocking material. The concern in this case is that most of the water bodies in which cage aquacultures are placed in Bulgaria at the moment are part of or on the boundaries of NATURA 2000 areas, and the invasion of exotic species is extremely undesirable. In contrast with Dospat Reservoir, in other investigated water bodies such as reservoirs with irrigation purposes and active sandpit lakes, the recorded abundance of pumpkinseed was relatively low. The observed low abundance might be explained by the fluctuating hydrological conditions in connection with irrigation and sand and gravel acquisition.

In most cases, the littoral zones of artificial and natural water bodies are strongly influenced by human activities. It is considered that habitat alteration by humans may be responsible for the establishment of some invasive fish species in Mediterranean-type streams and the increase in species invasiveness (Vila-Gispert et al. 2005, Almeida et al. 2009). Van Kleef et al. (2008) found that pumpkinseed is more abundant in fishing and moorland ponds, urban waters, and also water basins from which organic sediments, macrophytes, or inert materials have been obtained. Analysis of the relation between water body management and pumpkinseed abundance confirms the role of humans in the growth of pumpkinseed populations. The pumpkinseed abundance correlated positively with sandy and silty bottom substrate and this observation was in accordance with the reported pumpkinseed preference for constructing their nests more frequently on the sandy type of substrate (Danylchuk and Fox 1994). The other parameter that has a positive influence on pumpkinseed abundance is the density of submerged water vegetation. Without statistical conformation we observed that in the sites where the coastal vegetation is grass or reed, the species was more abundant. This is most likely due to the better exposure to the sun of sites surrounded by grass and numerous shelters in the reed belt. The lack of a statistically significant influence of predators on pumpkinseed abundance is not expected based on our observation that pumpkinseed was the most frequently found prey in the perch pike’s stomach (personal observation, unpublished).

It might be concluded that former sand-pit lakes give optimal opportunities for pumpkinseed breeding and growth. There are several prerequisites facilitating this process.

### Table 2
Effects of macro habitat parameters on the abundance, body size, and age of pumpkinseed, *Lepomis gibbosus* L. in Bulgaria

| Macrohabitat parameters | Abundance | Size | Age |
|------------------------|-----------|------|-----|
| Water body use         | df 4.26   | F 8.92 | P 0.0011<sup>a</sup> | 6.53 | 0.0008<sup>a</sup> | 2.85 | 0.04<sup>a</sup> |
| Surface area           | df 1.29   | F 1.09 | P 0.3 | 0.18 | 0.68 | 0.41 | 0.52 |
| Maximum depth          | df 1.29   | F 2.9 | P 0.09 | 0.41 | 0.52 | 0.12 | 0.73 |
| Altitude               | df 1.29   | F 0.18 | P 0.67 | 1.27 | 0.27 | 0.03 | 0.85 |
| Dissolve oxygen        | df 1.29   | F 3.26 | P 0.08 | 3.63 | 0.07 | 3.28 | 0.08 |
| Oxygen saturation      | df 1.29   | F 0.35 | P 0.55 | 3.34 | 0.08 | 0.82 | 0.19 |
| pH                     | df 1.29   | F 1.73 | P 0.2 | 1.51 | 0.23 | 1.55 | 0.22 |
| Temperature            | df 1.29   | F 4.84 | P 0.06 | 0.71 | 0.4 | 2.52 | 0.12 |
| Electro conductivity   | df 1.29   | F 0.08 | P 0.78 | 1.17 | 0.28 | 3.44 | 0.08 |

<sup>a, b, c</sup> Significant at *P* ≤ 0.05, 0.01, and 0.001, respectively; df = degrees of freedom.

### Table 3
Effects of microhabitat parameters on the abundance, body size, and age of pumpkinseed, *Lepomis gibbosus* L. in Bulgaria

| Microhabitat parameters | Abundance | Size | Age |
|------------------------|-----------|------|-----|
| Mean site depth        | df 1.29 | F 1.99 | P 0.17 | 0.73 | 0.39 | 0.06 | 0.8 |
| Silt [%]               | df 1.29 | F 18.56 | P 0.0001<sup>a</sup> | 9.11 | 0.005<sup>a</sup> | 10.12 | 0.003<sup>a</sup> |
| Sand [%]               | df 1.29 | F 18.47 | P 0.002<sup>a</sup> | 5.2 | 0.03<sup>a</sup> | 8.53 | 0.006<sup>a</sup> |
| Gravel [%]             | df 1.29 | F 0.28 | P 0.6 | 4.89 | 0.03<sup>a</sup> | 2.11 | 0.16 |
| Boulders [%]           | df 1.29 | F 0.64 | P 0.43 | 0.7 | 0.41 | 0.79 | 0.31 |
| Coastal vegetation     | df 2.28 | F 0.7 | P 0.5 | 0.66 | 0.53 | 0.73 | 0.49 |
| Aquatic vegetation     | df 3.27 | F 4.29 | P 0.01<sup>a</sup> | 3.28 | 0.03<sup>a</sup> | 1.78 | 0.17 |
| Predators              | df 1.29 | F 0.07 | P 0.93 | 8.09 | 0.008<sup>a</sup> | 10.41 | 0.003<sup>a</sup> |
| Bank slope             | df 1.29 | F 3.76 | P 0.06 | 3.87 | 0.037<sup>a</sup> | 1.18 | 0.28 |

<sup>a, b, c</sup> Significant at *P* ≤ 0.05, 0.01, and 0.001, respectively; df = degrees of freedom.
Firstly, the enlargement of sections with sandy bottoms in the littoral zones of excavation lakes led to an increase of suitable breeding habitats for pumpkinseed. Another significant factor is the removal of riparian vegetation along the coast of sand-pit lakes. Reduced shading of riparian zones led to elevation of the littoral temperatures, and while pumpkinseed is able to survive at temperatures over 30°C (Gutiérrez-Estrada et al. 2000), these temperatures are extremely disadvantageous for many other fishes. After stopping the exploitation of these lakes, fluctuation of the water level also stops. Additionally, most of these lakes become places for recreational fishing, which puts pressure on the population of native fish species. In this way, human activities facilitate the establishment of pumpkinseed populations, mainly by extending suitable breeding sites for pumpkinseed and eliminating its natural competitors and predators.

The observed fish diversity in the littoral zones of the investigated water bodies was relatively poor. Similar situations were observed in other European water bodies (García-Berthou and Moreno-Amich 2000). Nowadays, there are plenty of factors that could significantly reduce native fish species (unfavourable hydrochemical and hydrophysical conditions, overfishing, invaders, and others). There is no exact answer concerning the extent to which pumpkinseeds have influenced biodiversity and the abundance of native fish species, especially when there are no records from the time before the appearance of pumpkinseeds. We even found that with the growing abundance of pumpkinseeds the species number also increased. It is considered that the communities most vulnerable to *L. gibbosus* invasion are characterised by low species richness (Moyle and Light 1996, Rahel 2002).
Even if there is not sufficient contribution to the low abundance and biodiversity, the pumpkinseed managed to use the conditions and to occupy vacant niches and unused resources in the water ecosystem. The demographic structure of pumpkinseed-inhabited non-native zones of different types of water bodies is influenced by a variety of factors (Copp et al. 2004, Jordan et al. 2009). It is thought that the smaller size and age of the European pumpkinseed populations are a result of the influence of lower mean annual temperatures, limited food resources, and others (Villeneuve et al. 2005). The observed relatively low mean length might be explained by strong intrapopulation dietary overlap or dietary competition with other littoral species (Copp et al. 2004, Klaar et al. 2004). The smallest mean size of pumpkinseeds observed in highly polluted rehabilitation lakes might be due to strongly suppressed development of plankton and benthos. In contrast, in water bodies with a higher level of eutrophication and suitable conditions for the development of trophic resources such as reservoirs with aquaculture activities, pumpkinseeds have larger sizes. The existence of predators in the littoral zone decreases the mean age and size of pumpkinseeds in comparison with zones free of predators, but did not influence pumpkinseed abundance. By revealing the relations between some biological traits and environmental parameters, predictive models could be created to forecast the types of water bodies that are more vulnerable to invasion by alien species. Currently the experience of pest control over the pumpkinseed population is limited (van Klee et al. 2008). Potentially applicable measures for control could be activities such as manual destruction of the pumpkinseeds’ nests, restocking with predator species (e.g., perch and pike-perch), and a simultaneous reduction of the recreational fishing pressure on the native species. In small, shallow water bodies with lower biodiversity, drainage in the summer could be useful. Control over the process of eutrophication by limitation of biogenic elements flowing into the water body through some waste waters and nutrient inflow from agriculture and aquaculture has had a positive effect on the reduction of the populations. Some of these activities could be preventative, mainly in uninvaded water bodies. Water bodies inhabited by rare or endangered species should be more strictly controlled. Recreational fisheries should not be underestimated as a vector of introduction through the use of the pumpkinseed as a bait fish (Welcomme 1991, Elvira and Almodóvar 2001, Cambray 2003, Alcaraz et al. 2005, Marchetti et al. 2004, Moyle and Marchetti 2006). With regard to this, fishermen require more information in order to prevent new invasions.

CONCLUSIONS
The results of the study showed that the variation in the abundance and demographic structure of pumpkinseed corresponded with the type of substrate, density of underwater vegetation and use of the water body. This species finds optimal conditions in former sand-pit lakes and reservoirs where the cage aquaculture activities are well developed, bottom of the littoral zones are covered with sand and submerged vegetation is dense.

It can be concluded that human activities that leads to such alterations of littoral zones and habitat parameters, may be used in the elaboration of prediction models for the risk assessment of different types of water bodies and the potential invasiveness of pumpkinseed.

ACKNOWLEDGEMENTS
The authors wish to thank to two anonymous reviewers for their helpful comments on the manuscript. The authors would like also to thank Dr. Naidenov from the National Agency of Fisheries and Aquaculture for the assistance in fish sampling in Sopot Reservoir. The investigations were funded by National Scientific Fund (Ministry of Education, Youth and Science, Bulgaria), Grant N 01-0146 /2009-2011.

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Received: 14 June 2011
Accepted: 18 May 2012
Published electronically: 30 June 2012