Effect of straw return on soil respiration and NEE of paddy fields under water-saving irrigation

Shihong Yang¹,²*, Yanan Xiao², Junzeng Xu¹,², Xiaoyin Liu²

¹ State Key Laboratory of Hydrology-Water Resources and Hydraulic Engineering, Hohai University, Nanjing, P.R. China, ² College of Agricultural Engineering, Hohai University, Nanjing, P.R. China

* ysh7731@hhu.edu.cn

Abstract

Straw return (SR) and rice water-saving irrigation (WSI) affect the greenhouse gas emission of paddy fields. However, studies on CO₂ exchange between paddy fields and the atmosphere with joint regulation of SR and WSI are few. We conducted a two-year field experiment to investigate the effects of SR on soil respiration and net ecosystem exchange of CO₂ (NEE) in paddy fields under controlled irrigation (CI), which is a typical WSI technique. The rice yields, irrigation water use efficiency, seasonal variations in soil respiration, NEE, and soil organic carbon content were measured. Compared with the control (flooding irrigation and traditional chemical fertilizer), a significant increase in rice yield and irrigation water use efficiency in the paddy fields under CI and SR joint management (CS) was observed. CS increased the soil respiration rate during most of the rice growth stage and increased the net CO₂ absorption rate before approximately 80 days after transplanting; afterward, the pattern reversed. Total CO₂ emissions through soil respiration in CS paddy fields increased by 43.7% and 182% compared with the control in 2014 and 2015, respectively. However, CS also caused an increase in the total net CO₂ absorption by 18.1% and 30.1% in these two years, respectively. The acceleration in the consumption and decomposition of soil organic carbon induced by frequent alternate wet–dry cycles of the CI paddy fields increased the soil respiration and decreased the net CO₂ absorption. SR promoted soil respiration but also improved rice growth, increasing the net CO₂ absorption. The soil organic carbon content of the CS paddy fields after harvesting increased by 23.2% compared with that before transplanting. The present study concluded that joint regulation of WSI and SR is an effective measure for maintaining yield, increasing irrigation water use efficiency, mitigating CO₂ emission, and promoting paddy soil fertility.

Introduction

The increasing use of straw return (SR), an important management practice in global organic agriculture [1], is recommended to decrease chemical inputs, promote soil C sequestration, and improve crop yields [2–5]. However, SR has been shown to increase greenhouse gas...
For example, Zhang et al. [6] showed that SR increased the soil respiration rate by 9.60% in dry farmland. Liu et al. [7] found that SR increased methane (CH\textsubscript{4}) emissions by 111% in rice paddies and increased nitrous oxide (N\textsubscript{2}O) emissions by 90.0% in upland soils. Agricultural ecosystems are a major source of GHG emissions. The average annual total GHG emissions from agriculture reached 5.00–5.80 Gt CO\textsubscript{2} eq year\textsuperscript{−1} in 2000–2010, accounting for approximately 12.0% of total anthropogenic GHG emissions [8]. Thus, the environmental effects of SR application require comprehensive evaluation.

Rice is one of the most important cereal crops in the Asian monsoon region, and in 2014, the harvest area in this region was 143 million ha, accounting for 88% of the total rice harvest area worldwide [9]. With increasing water scarcity due to climate change, water-saving irrigation (WSI) techniques are being widely implemented in rice paddies [10,11]. The common point of these WSI techniques is non-flooding management or in unsaturated state of paddy soil during some or most of the rice growth season, leading to water conditions that are different from traditional flooding irrigation (FI). Thus, under such management, paddy fields experience multiple dry–wet cycles and consequently undergo changes in soil biological and chemical processes [12]. These results lead to changes that improve the effects of SR on GHG emissions from paddy fields. Zou et al. [13] found that under a water regime of flooding mid-season, drainage, reflooding, moist intermittent (F-D-F-M) irrigation, wheat straw and rapeseed cake incorporation resulted in a 252% increase in CH\textsubscript{4} emissions; moreover, rapeseed cake increased N\textsubscript{2}O by 17.0%, and wheat straw reduced N\textsubscript{2}O by 19.0% compared with controls. However, double rice-cropping system experiments indicated that midseason drainage and F-D-F-M irrigation reduced CH\textsubscript{4} emissions by 52.5% and 69.3% from paddy fields with SR, respectively, whereas F-D-F-M increased N\textsubscript{2}O emissions by 60.9% [14]. Existing studies on the effects of SR on GHG emission from paddy fields under WSI have focused mainly on CH\textsubscript{4} and N\textsubscript{2}O emissions [15–18].

Carbon dioxide (CO\textsubscript{2}) is another important GHG emitted from farmlands. When paddy soil is exposed to multiple wet–dry cycles under WSI, emission patterns and total emissions of CO\textsubscript{2} under SR will change. However, limited information on this change is available. The soil respiration and net ecosystem exchange of CO\textsubscript{2} (NEE) are two important characteristic values of CO\textsubscript{2} exchange between farmland and the atmosphere. Using the widely adopted WSI technique, we conducted a field experiment to identify the influence of SR on soil respiration and NEE of paddy fields under non-flooding controlled irrigation (CI) management. The objectives of this study were to (1) reveal the effects of water and carbon management on rice yield and irrigation water use efficiency; (2) analyze and compare the characteristics of seasonal variation in the soil respiration rate and the NEE of a paddy field ecosystem under different types of water and carbon management; (3) quantify total CO\textsubscript{2} emissions through soil respiration (total \(R_{soil}\)) and the total NEE of a paddy field ecosystem; and (4) discuss the effects of water and carbon management on soil respiration and the NEE of a paddy field ecosystem. The results can support more comprehensive evaluations of the ecological and environmental effects of rice WSI and SR. At the same time, they will also contribute to the comprehensive evaluation of GHG emissions and the sustainable use of water and carbon resources of paddy fields in China.

**Materials and methods**

**Site description**

The field experiment was conducted in 2014–2015 at the Kunshan Irrigation and Drainage Experiment Station in the Taihu Lake Region of China (31° 15′ 15″ N latitude, 120° 57′ 43″ E longitude). A rice–wheat rotation is used in this region. The paddy soil in the experimental site
is a clay-textured hydragric anthrosol (75.0% clay, 16.2% silt, and 8.80% sand) with 21.9 g kg\(^{-1}\) organic matter, 1.03 g kg\(^{-1}\) total nitrogen, 1.35 g kg\(^{-1}\) total phosphorus, 20.9 g kg\(^{-1}\) total potassium, and a pH of 7.40. In 2014 and 2015, the mean temperatures were 24.5˚C and 24.4˚C, and precipitations were 443 and 450 mm during the experimental period, respectively.

**Field management**

The experiment was laid out (plot size 150 m\(^2\)) in a randomized block design with four treatments and three replicates. The four treatments were a combination of irrigation and fertilizer managements: the two irrigation managements were CI and FI, and the two fertilizer managements were farmers’ fertilization practice (FFP) and wheat SR. Then, the four treatments were FF (FI and FFP), CF (CI and FFP), FS (FI and SR), and CS (CI and SR). All treatments were applied to the same plots for both years of the study. Rain-fed wheat was grown in the plots during the non-rice season. Non-flooding management was carried out in the CI treatment except for the shallow flooding water during the regreening stage. Moreover, irrigation was applied to saturate the soil only when the soil moisture approached the low threshold in a certain stage, as listed in Table 1. For the FI treatment, 3–5 cm of standing water was maintained in the paddy field except during mid-drainage in the late tillering stage. The soil moisture and water table were monitored at 8:00 a.m. every day throughout the rice growth stage.

Before the application of fertilizers, herbicides, and pesticides, all the plots were flooded to a depth of 3–5 cm. In addition, the same kinds and amounts of herbicides and pesticides were applied in both irrigation managements.

The rice variety used in this experiment was Japonica Rice Nanjing 46. Three to four seedlings per hill were transplanted with 13.0 cm × 25.0 cm hill spacing in late June and harvested in late October. Local nitrogen fertilizer was adopted in this experiment, as shown in Table 2. In addition to nitrogen fertilizer input, the same phosphorus and potassium fertilizers were applied to all treatments (56.3 kg P\(_2\)O\(_5\) ha\(^{-1}\) and 56.3 kg K\(_2\)O ha\(^{-1}\) in 2014, 54.0 kg P\(_2\)O\(_5\) ha\(^{-1}\) and 76.5 kg K\(_2\)O ha\(^{-1}\) in 2015). The chemical fertilizer management of the SR treatment was similar to that of the FFP treatment, and 3000 kg ha\(^{-1}\) of straw from the previous wheat crop (the organic carbon content of wheat straw was 441 g kg\(^{-1}\), and the organic carbon input through wheat straw was 1322 kg ha\(^{-1}\)) was returned to the SR paddy fields in both years. Base fertilizer and wheat straw were incorporated into the soil during tillage, and surface application was adopted for all other fertilizers.

**Field measurement and sampling**

Soil respiration and NEE of the paddy fields were measured using a transparent static chamber WEST Systems portable soil flux meter (West Systems S.r.l., Italy), which was described in

| Limit     | Regreening stage | Tillering stage | Jointing and booting stage | Heading and flowering stage | Milk stage | Ripening stage |
|-----------|------------------|-----------------|-----------------------------|-----------------------------|------------|----------------|
|           |                  | Initial | Middle | Late |                     |            | Naturally drying |
| Upper limit \(^2\) | 25 mm\(^1\) | 100% θ\(_{s1}\) | 100% θ\(_{s1}\) | 100% θ\(_{s1}\) | 100% θ\(_{s2}\) | 100% θ\(_{s3}\) | 100% θ\(_{s4}\) |
| Lower limit | 5 mm\(^1\) | 70% θ\(_{s1}\) | 65% θ\(_{s1}\) | 60% θ\(_{s1}\) | 75% θ\(_{s2}\) | 80% θ\(_{s3}\) | 70% θ\(_{s3}\) |
| Observed root zone depth (cm) | — | 0–20 | 0–20 | 0–20 | 0–30 | 0–40 | 0–40 |

\(^1\) Data show the water depth during the regreening stage. θ\(_{s1}\), θ\(_{s2}\), and θ\(_{s3}\) represent average volumetric soil moisture for the 0–20, 0–30, and 0–40 cm layers, respectively.

\(^2\) In the case of pesticide, fertilizer applications and rainfall, standing irrigation water at a depth of up to 5 cm is maintained for less than five days.

https://doi.org/10.1371/journal.pone.0204597.t001
During the rice growth stage, soil respiration and NEE were measured in every plot at 10:00–11:00 a.m. at 7-day intervals. Sampling bases with and without rice were used to measure the NEE and soil respiration fluxes of paddy fields, respectively. The cumulative CO$_2$ emissions through soil respiration (total $R_{\text{soil}}$) and NEE of the paddy field ecosystem during the study period were estimated by integrating emission fluxes across time.

Time domain reflectometer (Soil Moisture Equipment, Ltd., Corp. USA) and vertical rulers were used to monitor soil moisture and water depths, respectively. A water meter (Xiamen Longteng Industrial Co., Ltd., China) was installed on the pipe for each plot to measure the irrigation water volumes. The yield for each plot was measured after the rice ripened. Irrigation water use efficiency was calculated by dividing the yield by the irrigation water volume.

The soil samples were collected before transplanting and after harvesting in three replicates in each plot at 0–10, 10–20, and 20–40 cm depths. The soil samples for given depths were mixed and analyzed in the laboratory. Soil organic carbon content was determined using the potassium dichromate external heating method [20].

Statistical analyses were performed using standard procedures for a randomized plot design (SPSS 13.0, SPSS Inc., Chicago, IL). Significance was calculated using F-tests, and least significant differences were measured at the 0.05 probability level.

### Results

#### Rice yield and irrigation water use efficiency

Fifteen and thirteen wet–dry cycles occurred in the CI paddy fields in 2014 and 2015, respectively, representing more than 80 days of non-flooding conditions in both years (Figs 1 and 2). For the FI treatments, flooding was maintained, except during 42–44 and 49–51 days after transplantation (DAT) in 2014 and 34–45 DAT in 2015. These periods corresponded to the drainage in the late tillering stage to restrain nonproductive tillering.

No significant difference in rice yield was observed between the different irrigation treatments (Table 3). However, water input was dramatically reduced in the CI treatment by 49.5% and 43.3% in 2014 and 2015 ($p<0.05$), respectively, compared with FI paddy fields. Given the significant reduction in irrigation water input, CI treatment obviously improved the irrigation water use efficiency of the paddy fields. Irrigation water use efficiencies of the CI paddy fields with different carbon managements were increased by 96.4% and 75.8% compared with those of the FI fields in 2014 and 2015 ($p<0.05$), respectively.

SR significantly increased rice yield and irrigation water use efficiencies of paddy fields ($p<0.05$). Rice yield and irrigation water use efficiencies of the SR paddy fields under different irrigation treatments increased by 4.14%–7.91% compared with those of the FFP paddy fields. The interaction effect of irrigation and fertilizer treatments on rice yield was not significant.

### Table 2. Date and rate of nitrogen fertilization during the rice-growing season (kg N ha$^{-1}$).

| Activity                        | 2014                      | 2015                      |
|---------------------------------|---------------------------|----------------------------|
| Base fertilizer (19 and 29 Jun) | 159(56.3CF+103AB)         | 155(72.0CF+83.2AB)        |
| Tillering fertilizer (29 Jun and 5 Jul) | 76.2(U)                 | 69.3(U)                   |
| Panicle fertilizer (10 and 9 Aug) | 55.4(U)                   | 58.9(U)                   |
| Total nitrogen                  | 291                       | 283                       |

Date in the bracket is the time for the fertilizer applied in 2014 and 2015 respectively.

CF: compound fertilizer (N, P$_2$O$_5$ and K$_2$O contents are 15.0%, 15.0% and 15.0% in 2014, and 16.0%, 12.0% and 17.0% in 2015), AB: ammonium bicarbonate (N content is 17.1%), U: urea (N content is 46.2%).

https://doi.org/10.1371/journal.pone.0204597.t002
but was significant for irrigation water use efficiency (Table 4). In addition to a significant reduction in irrigation water input, the joint regulation of WSI and SR also increased the rice yields and irrigation water use efficiencies compared with traditional irrigation and fertilizer management. Compared with FF paddy fields, the rice yields of the CS paddy fields increased by 7.14% and 3.77% in 2014 and 2015 (p<0.05), respectively. Irrigation water use efficiencies of the CS paddy fields were 2.12 and 1.83 times higher than those of FF paddy fields in 2014 and 2015, respectively.

Seasonal variations in soil respiration

Soil respiration rates exhibited an upward trend after transplanting and peaked during late July or August. Then, they fluctuated but generally exhibited a downward trend until the late milk stage (Fig 3). During the ripening stage, soil respiration rates showed a significant increase during the natural drying period.

The soil respiration enhancement effects of WSI differed between different fertilizer treatments. Before the ripening stage (before 110 DAT), the soil respiration rates of the CI paddy
fields with FFP were mostly larger than those of the FI paddy fields. For the SR paddy fields, soil respiration rates of the CI paddy fields were only larger than those of the FI paddy fields.

Table 3. Rice yield and irrigation water use efficiency.

| Items                              | FF            | FS            | CF            | CS            |
|------------------------------------|---------------|---------------|---------------|---------------|
| 2014                               |               |               |               |               |
| Yield (kg ha⁻¹)                    | 9657±54.4b    | 10422±132a    | 9589±88.2b    | 10347±70.8a   |
| Irrigation water volume (mm)       | 804±11.0a     | 804±11.0a     | 407±5.50b     | 407±5.50b     |
| IWUE (kg m⁻³)                      | 1.20±0.0104d  | 1.30±0.00702c | 2.36±0.0241b  | 2.55±0.0190a  |
| 2015                               |               |               |               |               |
| Yield (kg ha⁻¹)                    | 9730±40.3b    | 10119±112a    | 9682±61.8b    | 10096±80.4a   |
| Irrigation water volume (mm)       | 912±11.9a     | 912±11.9a     | 517±5.73b     | 517±5.73b     |
| IWUE (kg m⁻³)                      | 1.07±0.0964d  | 1.11±0.00418c | 1.87±0.00875b | 1.95±0.00695a |

FF: flooding irrigation and farmers’ fertilization practice, FS: flooding irrigation and wheat straw return at a rate of 3000 kg ha⁻¹, CF: controlled irrigation and farmers’ fertilization practice, CS: controlled irrigation and wheat straw return at a rate of 3000 kg ha⁻¹, IWUE: irrigation water use efficiency. Means in the same line in 2014 or 2015 followed by the same letter are not significantly different (p < 0.05).

https://doi.org/10.1371/journal.pone.0204597.t003
Table 4. MANOVA results for rice yield and irrigation water use efficiency.

| Year | Influence factor       | Rice yield | Irrigation water use efficiency |
|------|------------------------|------------|---------------------------------|
|      |                        | SS  | F    | P      | SS  | F    | P      |
| 2014 | Fertilizer management  | 1.73×10^6 | 69.9 | 3.18×10^{-5} & 5.94×10^{-2} | 71.7 | 2.89×10^{-5} |
|      | Water management       | 1.54×10^4 | 0.617 | 0.455 | 4.35 | 5.25×10^{-1} | 1.47×10^{-12} |
|      | Interactive effect     | 33.2 | 1.33×10^{-3} | 0.972 | 6.30×10^{-3} | 7.60 | 2.48×10^{-2} |
|      | Error                  | 1.99×10^5 | 6.63×10^{-3} | | |
| 2015 | Fertilizer management  | 4.84×10^4 | 26.4 | 8.89×10^{-4} | 1.13×10^{-2} | 63.9 | 4.40×10^{-5} |
|      | Water management       | 3.70×10^3 | 0.202 | 0.665 | 2.04 | 1.16×10^1 | 6.23×10^{-14} |
|      | Interactive effect     | 478 | 2.60×10^{-2} | 0.876 | 1.06×10^{-3} | 6.02 | 3.98×10^{-11} |
|      | Error                  | 1.47×10^5 | 1.41×10^{-3} | | |

SS: sum of squares of mean deviation.
*: significant at 0.05 level

https://doi.org/10.1371/journal.pone.0204597.t004

before the beginning of late August (approximately 55 DAT). The values of soil respiration rates under different irrigation treatments also crossed each other before the ripening stage. However, the soil respiration rates of the FI paddy fields were greater than those of the CI.

Fig 3. Soil respiration rates of paddy fields with different water and carbon managements FF: Flooding irrigation and farmers’ fertilization practice, FS: Flooding irrigation and wheat straw return at a rate of 3000 kg ha⁻¹, CF: Controlled irrigation and farmers’ fertilization practice, CS: Controlled irrigation and wheat straw return at a rate of 3000 kg ha⁻¹, DAT: Day after transplantation.

https://doi.org/10.1371/journal.pone.0204597.g003
paddy fields during the ripening stage regardless of fertilizer treatment. For the FFP paddy fields, the average soil respiration rates of the CI paddy fields were 0.438 and 0.267 mol m$^{-2}$ day$^{-1}$ in 2014 and 2015, representing increases of 27.9% and 35.7% compared with those of the FI paddy fields, respectively. However, for the SR paddy fields, the increase in WSI on soil respiration diminished. In addition, the average soil respiration rates of the CI paddy fields increased by 10.9% and 5.23% compared with those in the FI paddy fields in 2014 and 2015, respectively. This interannual variation in soil respiration under the WSI treatment may be attributed to the following factors. Irrigation treatment was the main factor that influenced soil respiration in 2014 due to the similar soil respiration rates of paddy fields under the same irrigation treatments during the rice growth stage. In addition, the soil respiration peaks of paddy fields under the same irrigation treatments occurred at the same time in both years of the study. However, SR had a greater effect on soil respiration in the second year.

SR increased the soil respiration rates of the paddy fields, but the effect differed across irrigation treatments and years. For the CI paddy fields, soil respiration rates of the SR paddy fields were consistently higher than those of the FFP paddy fields before the middle of August (approximately 50 DAT) in 2014. Then, the values of soil respiration rates with different fertilizer managements converged. In the same year, the soil respiration rates of the FS paddy fields were higher than those of the FF paddy fields across most growth stages. In 2015, the soil respiration rates of the SR paddy fields were significantly higher than those of the FFP paddy fields under different water managements ($p<0.05$). There was an interannual difference in the increase in soil respiration associated with SR. In 2014, the average soil respiration rates of the FS and CS paddy fields were 0.413 and 0.458 mol m$^{-2}$ day$^{-1}$, respectively. Moreover, these values were 20.5% and 4.49% higher than those of the FF and CF paddy fields, respectively. The enhanced effect of SR on soil respiration in the second year was obviously higher than that in the first year. The average soil respiration rates of FS and CS paddy fields were 0.644 and 0.678 mol m$^{-2}$ day$^{-1}$ and were 3.27 and 2.53 times higher in 2014 than those of FF and CF paddy fields in 2015, respectively. In addition, soil respiration peaks of fields under the SR treatments occurred on the same day. The maximum soil respiration rates of the FS and CS paddy fields both occurred at 42 DAT. Moreover, the peak values were 1.80 and 2.23 mol m$^{-2}$ day$^{-1}$ in 2015, respectively.

Average soil respiration rates of the CS paddy fields were 0.457 and 0.678 mol m$^{-2}$ day$^{-1}$ in 2014 and 2015, showing an increase of 33.6% and 244%, respectively, compared with those of FF paddy fields.

**Seasonal variations in NEE**

During the early growth stage, the net CO$_2$ absorption rate of the paddy field ecosystem increased with rice growth (Fig 4) and fluctuated after the peak at approximately 60–70 DAT in 2014 and 30–50 DAT in 2015. The net CO$_2$ absorption rate significantly decreased with the decrease of the assimilation ability during the rice maturation stage.

The ecosystem NEE showed similar variations under different irrigation treatments during the early rice growth stage (before approximately 45 and 30 DAT in 2014 and 2015, respectively). Thereafter, the non-flooding periods of the WSI treatment resulted in lower net CO$_2$ absorption rates of the CI paddy field ecosystems than those of the FI paddy fields under the FFP treatment until the ripening stage (107–123 DAT in 2014 and 106–121 DAT in 2015). However, the values of NEE for SR paddy fields under different irrigation treatments crossed each other until the ripening stage. During the ripening stage, the variation in the net CO$_2$ absorption rates contrasted with the soil respiration rate. Moreover, the low net CO$_2$ absorption rates for the FI paddy fields were lower than those of the CI paddy fields. For the FFP
paddy fields, the average net CO$_2$ absorption rates of the CF paddy fields decreased by 15.4% and 13.5% compared with those of the FF paddy fields in 2014 and 2015, respectively. For the SR paddy fields, the average net CO$_2$ absorption rates of the CS paddy fields increased by 3.58% and 18.7% compared with those of the FS paddy fields in 2014 and 2015, respectively. Overall, SR increased the net CO$_2$ absorption rates of the paddy fields. However, its influence differed between different irrigation treatments. For the FI paddy fields, net CO$_2$ absorption rates of the FS paddy fields were larger than those of the FF paddy fields most of the time before approximately 70 DAT. Then, the pattern reversed. For the CI paddy fields, the net CO$_2$ absorption rates of the CS paddy fields were higher than those of the CF paddy fields most of time before the ripening stage. The net CO$_2$ absorption rates of the CS paddy fields decreased relative to those of CF paddy fields during the ripening stage. The average net CO$_2$ absorption rates of the FS paddy fields were 0.952 and 0.842 mol m$^{-2}$ day$^{-1}$ in 2014 and 2015, indicating an 11.58% and 21.52% increase, respectively, compared with the FF paddy fields. The average net CO$_2$ absorption rates of the CS paddy fields represented a significant increase of 36.7% and 66.7% compared with the CF paddy fields in 2014 and 2015 ($p<0.05$), respectively.

Compared with traditional water and fertilizer management, the combination of WSI and SR increased the net CO$_2$ absorption rates before approximately 80 DAT. Afterward, the pattern reversed, and the average net CO$_2$ absorption rates increased by 15.6% and 42.2% in 2014 and 2015, respectively.
Total $R_{soil}$ and NEE throughout the rice growth stage

Table 5 shows the total CO$_2$ emissions through soil respiration (total $R_{soil}$) and total NEE of the paddy field system. Under the FFP fertilizer treatment, WSI caused an increase in total $R_{soil}$ and a decrease in the total net CO$_2$ absorption compared with the FI paddy fields. Total $R_{soil}$ of the CF paddy fields increased by 15.2% and 8.16% compared with that of the FF paddy fields in 2014 and 2015, respectively. Total net CO$_2$ absorption of the CF paddy fields decreased by 11.8% and 11.3%, respectively, compared with that of the FF paddy fields. Under the SR fertilizer treatment, WSI caused an increase in total net CO$_2$ absorption ($p<0.05$) compared with FI. Total $R_{soil}$ of the CS paddy fields increased by 3.88% and 3.95% compared with the FF paddy fields in 2014 and 2015, respectively. Total net CO$_2$ absorption of the CS paddy fields increased by 9.73% and 13.4% compared with that of the FS paddy fields in 2014 and 2015, respectively.

SR caused a significant increase in the total $R_{soil}$ and net CO$_2$ absorption compared with FFP management ($p<0.05$), except for the increase in total net CO$_2$ absorption in 2014. Total $R_{soil}$ of SR paddy fields under different irrigation treatments increased by 31.6% and 166% in 2014 and 2015, respectively, compared with that of the FF paddy fields. The total net CO$_2$ absorption of SR paddy fields increased by 21.5% and 31.6% in 2014 and 2015, respectively, compared with that of the FF paddy fields.

The combination of WSI and SR resulted in a significant increase in the total $R_{soil}$ and total net CO$_2$ absorption compared with traditional irrigation and fertilizer treatments ($p<0.05$). Total $R_{soil}$ of the CS paddy fields was 75.9 and 110 mol m$^{-2}$ in 2014 and 2015, an increase of 43.8% and 182%, respectively, compared with that of the FF paddy fields. Total net CO$_2$ absorption of the CS paddy fields increased by 18.1% and 30.1%, respectively, compared with that of the FF paddy fields.

Table 6 shows the multivariate analysis of variance (MANOVA) results for total $R_{soil}$ and NEE. The results indicated that the fertilizer treatment had a significant effect on total $R_{soil}$ and NEE. The effect of fertilizer treatment accounted for 78.7% and 98.9% of the mean variance in the sum of squares (SS) for total $R_{soil}$ in 2014 and 2015 and 68.7% and 80.0% of the total variance for total NEE in 2014 and 2015, respectively. The interaction of water and fertilizer treatments had a significant effect on total NEE. In addition, its variance contributions were 25.2% and 17.2% of the SS for the total NEE in 2014 and 2015, respectively. The effect of water management on total $R_{soil}$ and NEE was not significant.

Soil organic carbon content

SR increased the postharvest soil carbon content relative to the pretransplant soil carbon content (Fig 5). Under the FS treatment, postharvest soil organic carbon content increased by

Table 5. Total $R_{soil}$ and NEE during the whole rice growth stage (mol m$^{-2}$).

| Year | Item | FF          | FS           | CF           | CS           |
|------|------|-------------|--------------|--------------|--------------|
| 2014 | $R_{soil}$ | 52.8±1.16c | 73.0±2.87a   | 60.8±2.62b   | 75.9±2.79a   |
|      | NEE  | -115±2.37bc | -123.2±2.36b | -99.9±1.89d  | -135±2.45a   |
| 2015 | $R_{soil}$ | 38.9±1.08b | 105.6±3.18a  | 42.0±0.973b  | 110±2.33a    |
|      | NEE  | -86.4±0.931c| -99.8±1.57b  | -76.6±0.39d  | -113±1.98a   |

FF: flooding irrigation and farmers’ fertilization practice, FS: flooding irrigation and wheat straw return at a rate of 3000 kg ha$^{-1}$, CF: controlled irrigation and farmers’ fertilization practice, CS: controlled irrigation and wheat straw return at a rate of 3000 kg ha$^{-1}$, $R_{soil}$: CO$_2$ emission through soil respiration, NEE: net CO$_2$ exchange between paddy fields ecosystem and atmosphere. Means in the same line in 2014 or 2015 followed by the same letter are not significantly different ($p<0.05$).

https://doi.org/10.1371/journal.pone.0204597.t005
35.3%, 32.7%, and 54.8% in 0–10, 10–20, and 20–40 depths, respectively, compared with the content prior to transplanting in 2015. The rate of increase under the CS treatment was relatively low. The rates increased by 17.1%, 5.76%, and 46.7% for the three depths, respectively. SR is recommended as the most effective and economic method of soil carbon sequestration in paddy field systems [21]. SR increases carbon input, which favors fungal growth. Adhesive organic molecules associated with particulate organic matter stabilization are produced during microbial-mediated straw decomposition [22].

Pure chemical fertilizer treatment reduced postharvest soil organic carbon content relative to pretransplant content. Accelerated decomposition of soil organic carbon caused by WSI reduced soil organic carbon content after harvesting relative to the FI paddy. Xu et al. found that rain-fed fields with irrigation applied only during drought periods exhibited significantly lower total soil organic carbon stock relative to FI [23]. Non-flooding management under the CI treatment had a similar effect on soil organic carbon content to that of rain-fed management.

Table 6. MANOVA results for total $R_{soil}$ and NEE.

| Year | Influence factor | $R_{soil}$ | NEE |
|------|------------------|------------|-----|
|      |                  | SS | $F$  | $P$ | SS | $F$  | $P$ |
| 2014 | Fertilizer management | 935 | 51.4 | 9.51x10^{-5} | 1.46x10^2 | 93.5 | 1.09x10^{-5}  |
|      | Water management | 88.0 | 4.84 | 5.91x10^{-2} | 5.54 | 0.355 | 0.568 |
|      | Interactive effect | 20.0 | 1.10 | 0.325 | 535 | 34.3 | 3.78x10^{-4} |
|      | Error | 146 | 125 | 0  | 0 |
| 2015 | Fertilizer management | 1.36x10^4 | 1.03x10^3 | 9.88x10^{-10} | 1.87x10^7 | 271 | 1.87x10^{-7} |
|      | Water management | 40.4 | 3.05 | 0.119 | 10.0 | 1.45 | 0.263 |
|      | Interactive effect | 0.753 | 5.69x10^{-2} | 0.818 | 402 | 58.3 | 6.09x10^{-5} |
|      | Error | 105.9 | 55.2 | 0  | 0 |

$R_{soil}$: CO$_2$ emission through soil respiration, NEE: net CO$_2$ exchange between paddy fields ecosystem and atmosphere, SS: sum of squares of mean deviation.  
*: significant at 0.05 level

Fig 5. Soil organic carbon content of paddy fields before transplanting and after harvesting FF: Flooding irrigation and farmers’ fertilization practice, FS: Flooding irrigation and wheat straw return at a rate of 3000 kg ha$^{-1}$, CF: Controlled irrigation and farmers’ fertilization practice, CS: Controlled irrigation and wheat straw return at a rate of 3000 kg ha$^{-1}$, BV: Background value.

https://doi.org/10.1371/journal.pone.0204597.g005
Discussion

Response of soil respiration to soil moisture and air temperature

An exponential function can used to describe the relationship between soil moisture and air temperature (Table 7). However, all the values of the coefficient of determination ($R^2$) were not high. This result may be due to the influence of numerous unmeasured factors on soil respiration under field conditions. Generally, the paddy soil respiration increased exponentially with the increase in air temperature. The relationship between soil respiration and air temperature varied in different treatments. Variation in the temperature sensitivity coefficient ($Q_{10}$) of the paddy soil respiration reveals differences between the treatments. $Q_{10}$ values were higher under the CI treatment relative to the FI treatment as the soil respiration rate is sensitive to air temperature in the CI treatment due to the absence of an insulating water layer over the soil. Fertilizer treatment did not significantly affect $Q_{10}$ values.

Soil moisture is another important factor that affects soil respiration. Soil respiration is generally believed to increase with the increase in soil moisture up to field moisture capacity. Then, the activity of aerobic microorganisms decreases under the resulting anaerobic conditions, and soil respiration decreases [24]. In this study, the water content of the WSI paddy field, which was higher than the field moisture capacity, resulted in a linear decrease in the soil respiration rate with the increase in soil water moisture (Fig 6).

Effect of SR on soil respiration and NEE of the paddy fields

In this experiment, SR increased the paddy soil respiration and improved the net CO$_2$ absorption of the paddy ecosystems. Most previous studies on the effects of straw addition on soil respiration and NEE were limited to dry farmland; this work showed that SR enhanced the soil respiration and net CO$_2$ absorption in dry farmland ecosystems [25–27]. Our results showed that straw addition has the same effect on the soil respiration and NEE of paddy fields as in dry farmland.

The reasons for the increases in paddy soil respiration due to SR may be as follows: (1) SR increased soil porosity and CO$_2$ concentration in soil solutions [28,29], and these changes facilitate the diffusion of CO$_2$ from paddy soil to the atmosphere and increase CO$_2$ emission. (2) SR increased soil organic matter content, thereby increasing soil CO$_2$ emissions [5,30]. Existing research has shown that soil organic carbon content and soil respiration are significantly positively correlated [31,32]. In this experiment, the soil organic carbon content was higher postharvest in the SR paddy fields than that prior to transplanting (Fig 5). (3) SR increases some available soil nutrients including phosphorus, potassium, organic carbon, and alkali-hydrolyzable nitrogen [33,34]. The changes in available soil nutrients affect the rate of carbon cycling, thereby influencing soil CO$_2$ emission. (4) SR affects soil microbial biomass.

Table 7. The relationship between soil respiration rate and air temperature.

| Treatment | Fitting equation | $R^2$ | $P$ | $n$ | $Q_{10}$ |
|-----------|------------------|-------|-----|-----|---------|
| FI        | $SR = 0.0029\exp(0.1289\times T)$ | 0.148 | <0.05 | 32 | 3.66    |
| CI        | $SR = 0.0036\exp(0.1404\times T)$ | 0.147 | <0.05 | 32 | 4.07    |
| FFP       | $SR = 0.0028\exp(0.1373\times T)$ | 0.142 | <0.05 | 32 | 3.95    |
| SR        | $SR = 0.0032\exp(0.1399\times T)$ | 0.136 | <0.05 | 36 | 3.90    |

SR: soil respiration rate, $T$: air temperature, $Q_{10}$: The temperature sensitivity coefficient of soil respiration, FI: treatments with flooding irrigation, contains FF and FS, CI: treatments with controlled irrigation, contains CF and CS, FFP: treatments with farmers’ fertilization practice, contains FF and CF, SR: treatments with wheat straw return, contains FS and CS.

https://doi.org/10.1371/journal.pone.0204597.t007
microbial community structure, and soil enzyme activity, promoting the metabolic activity of microbes and improving the soil respiration rate [35,36]. Meanwhile, SR promotes crop growth, yield, and dry biomass [30,37]. This increased crop growth also increased the CO$_2$ absorption through rice photosynthesis. Thus, the net CO$_2$ absorption of the paddy ecosystem increases along with the increased soil respiration rates in paddy fields under SR management relative to paddy fields without SR.

The increased respiration and net CO$_2$ absorption of paddy soil under SR exhibited significant interannual variation in this experiment (Table 5). The effects of SR in the second year (2015) were greater than that in the first year (2014), particularly for paddy soil respiration values. This phenomenon can be attributed to the slow decomposition rate of straw [38,39]. The reductive conditions of paddy fields were also an obstacle to straw decomposition compared with that of dry farmlands. This restricted the effect of SR on promoting paddy soil respiration and net CO$_2$ absorption for the first year. Dry farmland management during the rain-fed wheat stage (all the plots had the same field management during the rain-fed wheat stage) accelerated wheat straw decomposition and supplied reaction substrate for soil microorganisms and animals as well as nutrients for rice growth during the next season. This process resulted in the increased effect of straw on paddy soil respiration and net CO$_2$ absorption, which became more apparent in the second year. In addition, continuous application of wheat straw for two years may also help to explain this interannual variation. Previous research has shown that cotton soil respiration increases with the increase in SR over years of continuous cropping, and over 30 years of continuous SR increases cumulative CO$_2$ emission through soil respiration by 4.26% compared with that in the fifth year [40].

**Effects of water and straw management on rice production and CO$_2$ exchange of paddy fields**

A previous study has shown that various WSI management modes in China can reduce the volume of irrigation by 8%–50% and increase rice yield and irrigation water use efficiency by 3.00%–8.00% and 20.0%–80.0%, respectively [41]. A meta-analysis showed that alternate wetting and drying (AWD) decreased rice yields by 5.40% and irrigation water input by 25.7% compared with continuous flooding in Asia; however, under mild AWD, rice yields were not
significantly reduced in most circumstances. As a result, AWD increased irrigation water use efficiency by 24.2% [42]. WSI caused a slight decrease in rice yield in this study. In addition, WSI decreased the net CO₂ absorption of a paddy field ecosystem and soil organic carbon content by 9.73%–13.4% and 3.24%–20.3%, respectively, compared with FI (Table 5 and Fig 5). As a farmland management technology widely recommended in China, SR can decrease chemical inputs, promote soil C sequestration, and improve crop yields [3,4]. This research determined whether SR can resolve decreased rice yield and decreased net ecosystem CO₂ absorption of paddy fields under WSI. The results of this study showed that the joint regulation of WSI and SR reduced the irrigation water input by 43.3%–49.5% and increased rice yield by 3.77%–7.14% compared with traditional water and fertilizer management practices. Meanwhile, total net CO₂ absorption and soil organic carbon content increased by 18.1%–30.1% and 5.76%–46.7%, respectively. Therefore, the joint regulation of WSI and SR is an effective measure for maintaining yield, increasing irrigation water use efficiency, mitigating CO₂ emission, and promoting paddy soil fertility. In addition to CO₂, paddy fields are an important emission source of CH₄ and N₂O. Generally, SR markedly increases CH₄ and N₂O emissions from paddy fields under FI [4,43]. WSI can reduce CH₄ emission and increase N₂O emission from paddy fields [44,45]. However, few studies have focused on the effect of SR on CH₄ and N₂O emissions from paddy fields under WSI. WSI and SR techniques have been widely applied in paddy fields in China. Therefore, the greenhouse effects (CH₄, N₂O, and CO₂) of paddy fields with joint regulation of SR and WSI should be studied in depth.

The excessive nitrogen fertilizer input into paddy fields in China is also an urgent issue. Nitrogen fertilizer input during the rice season in the Taihu Lake region, which is one of the main rice-producing areas of China, reaches up to 270–300 kg N ha⁻¹ [46]. In this experiment, nitrogen fertilizer inputs, according to local conventional fertilizer application during the rice season, were 291 and 283 kg N ha⁻¹ in 2014 and 2015, respectively. However, the recommended amount of nitrogen fertilizer input is only 190–200 kg N ha⁻¹ [47]. In addition, the nitrogen use efficiency of paddy fields was only approximately 30.0% [48]. Much of the applied nitrogen fertilizer is lost through runoff, leaching, and ammonia volatilization. High rates of nitrogen fertilizer input and the loss of paddy fields in this region are contributors to the eutrophication of lakes and rivers. Therefore, measures such as introducing N fertilizer tax, improving local extension services, and educating farmers for environmental awareness should be taken to avoid excessive nitrogen fertilizer input and serious environmental degradation in the Taihu Lake region.

Conclusions

Understanding the effects of WSI combined with SR on the CO₂ exchange between a paddy field ecosystem and the atmosphere can support comprehensive evaluations of greenhouse effects and the sustainable use of the water and carbon resources of paddy fields. CS management significantly increased the rice yields and the irrigation water use efficiency of paddy fields compared with the control. CS clearly increased soil respiration rates during most of the rice growing season and increased net CO₂ absorption rates before approximately 80 DAT. Afterward, the pattern reversed. Total CO₂ emission through soil respiration of CS paddy fields increased by 43.7% and 182% compared with the control in 2014 and 2015, respectively. However, CS also raised the total net CO₂ absorption by 18.1% and 30.1% in these two years, respectively. Frequent alternating wet–dry cycles of the CI paddy fields led to an increase in soil respiration and a decrease in net CO₂ absorption. SR promoted paddy soil respiration but also increased the net CO₂ absorption and paddy soil organic carbon content. The present study concludes that the joint regulation of WSI and SR is an effective measure for maintaining
yield, increasing irrigation water use efficiency, mitigating CO$_2$ emission, and promoting paddy soil fertility.

**Supporting information**

S1 Data File. We provided the data used in this research. The caption “Fig 2” mentioned at Fig 2 as an example.

(XLSX)

**Acknowledgments**

This research was financially supported by the National Natural Science Foundation of China (No. 51579070, 51879076), the Fundamental Research Funds for the Central Universities (No. 2018B634X14), the Postgraduate Research & Practice Innovation Program of Jiangsu Province (KYCX18_0628) and the Advanced Science and Technology Innovation Team in Colleges and Universities in Jiangsu Province.

**Author Contributions**

Data curation: Shihong Yang, Yanan Xiao, Xiaoyin Liu.

Funding acquisition: Shihong Yang.

Methodology: Shihong Yang, Yanan Xiao, Junzeng Xu.

Writing – original draft: Shihong Yang.

**References**

1. Seufert V, Ramankutty N, Foley JA. Comparing the yields of organic and conventional agriculture. Nature. 2012; 485:229–232. [https://doi.org/10.1038/nature11069](https://doi.org/10.1038/nature11069) PMID: 22535250

2. Baudron F, Jaleta M, Okitoi O, Tegegn A. Conservation agriculture in African mixed crop-livestock systems: expanding the niche. Agr Ecosyst Environ. 2014; 187:171–182. [https://doi.org/10.1016/j.agee.2013.08.020](https://doi.org/10.1016/j.agee.2013.08.020)

3. Turmel MS, Speratti A, Baudron F, Verhulst N, Govaerts B. Crop residue management and soil health: a systems analysis. Agr Syst. 2014; 134:6–16. [https://doi.org/10.1016/j.agsy.2014.05.009](https://doi.org/10.1016/j.agsy.2014.05.009)

4. Hu NJ, Wang BJ, Gu ZH, Tao BR, Zhang ZW, Hu SJ, et al. Effects of different straw returning modes on greenhouse gas emissions and crop yields in a rice-wheat rotation system. Agr Ecosyst Environ. 2016; 223:115–122. [https://doi.org/10.1016/j.agee.2016.02.027](https://doi.org/10.1016/j.agee.2016.02.027)

5. Chen ZM, Wang HY, Liu XW, Zhao XL, Lu DJ, Zhou JM, et al. Changes in soil microbial community and organic carbon fractions under short-term straw return in a rice-wheat cropping system. Soil Tillage Res. 2017; 165:121–127. [https://doi.org/10.1016/j.still.2016.07.018](https://doi.org/10.1016/j.still.2016.07.018)

6. Zhang QZ, Wu WL, Wang MX, Zhou ZR, Chen SF. The effects of crop residue amendment and N rate on soil respiration. Acta Ecol Sin. 2005; 25(11):2883–2887.

7. Liu C, Lu M, Cui J, Li B, Fang CM. Effects of straw carbon input on carbon dynamics in agricultural soils: a meta-analysis. Global Change Biol. 2014; 20:1366–1381. [https://doi.org/10.1111/gcb.12517](https://doi.org/10.1111/gcb.12517) PMID: 24395454

8. IPCC (2014) Climate change 2014 mitigation of climate change. Working group III contribution to the fifth assessment report of the intergovernmental panel on climate change. Cambridge University Press.

9. FAO. Faostat. [http://www.fao.org/faostat/en/#data/QC](http://www.fao.org/faostat/en/#data/QC). 2015.

10. Mao Z. Water efficient irrigation and environmentally sustainable irrigated rice production in China. International Commission on Irrigation and Drainage Web. [http://www.icid.org/wat_mao.pdf](http://www.icid.org/wat_mao.pdf). 2001.

11. Bouman BAM, Lampayan RM, and Tuong TP. Water Management in Irrigated Rice: Coping with Water Scarcity. International Rice Research Institute (IRRI). Los Baños, Philippines. 2007.

12. Mao Z. Water saving irrigation for rice and its effect on environment. Eng Sci. 2002; (4): 8–16.
13. Zou JW, Huang Y, Jiang JY, Zheng XH, Sass RL. A 3-year field measurement of methane and nitrous oxide emissions from rice paddies in China: Effects of water regime, crop residue, and fertilizer application. Global Biogeochem Cycles. 2005; 19(2): GB2021. https://doi.org/10.1029/2004GB002401

14. Shang QY, Yang XX, Cheng C, Luo K, Huang S, Shi QH, et al. Effects of water regime on yield-scaled global warming potential under double rice-cropping system with straw returning. Chin J Rice Sci. 2015; 29(2):181–190.

15. Yan XY, Yagi K, Akiyama H, Akimoto H. Statistical analysis of major variables controlling methane emission from rice fields. Global Change Biol. 2005; 11(7):1131–1141. https://doi.org/10.1111/j.1365-2486.2005.00976.x

16. Ali MA, Hoque MA, Kim PJ. Mitigating global warming potentials of methane and nitrous oxide gases from rice paddies under different irrigation regimes. Ambio. 2013; 42(3):357–368. https://doi.org/10.1007/s13280-012-0349-3 PMID: 23015326

17. Malayan SK, Bhatia A, Kumar A, Gupta DK, Singh R, Kumar SS, et al. Methane production, oxidation and mitigation: A mechanistic understanding and comprehensive evaluation of influencing factors. Sci Total Environ. 2016; 572(1):874–896. https://doi.org/10.1016/j.scitotenv.2016.07.182 PMID: 2775427

18. Gupta DK, Bhatia A, Kumar A, Das TK, Jain N, Tomer R, et al. Mitigation of greenhouse gas emission from rice–wheat system of the Indo-Gangetic plains: Through tillage, irrigation and fertilizer management. Agr Ecosyst Environ. 2016; 230:1–9. https://doi.org/10.1016/j.agee.2016.05.023

19. Yu Y, Xu JZ, Liu XY, Zhang JG, Wang YJ. Variations of carbon dioxide exchange in paddy field ecosystem under water-saving irrigation in Southeast China. Agr Water Manage. 2016; 166:42–52. https://doi.org/10.1016/j.agwat.2015.12.015

20. Bao ST. Soil and agro-chemistry analysis. China agricultural press, Beijing, 1999; pp 264–268.

21. Li H, Dai MW, Dai SL, Dong XJ. Current status and environment impact of direct straw return in China’s cropland-A review. Ecotox Environ Saf. 2018; 159:293–300. https://doi.org/10.1016/j.ecoenv.2018.05.014 PMID: 29763811

22. Kunlanit B, Vityakon P, Puttaso A, Cadisch G, Rasche F. Mechanisms controlling soil organic carbon composition pertaining to microbial decomposition of biochemically contrasting organic residues: evidence from midDRIFTS peak area analysis. Soil Biol. Biochem. 2014; 76:100–108. https://doi.org/10.1016/j.soilbio.2014.05.006.

23. Xu Y, Zhan M, Cao CG, Ge JZ, Ye RZ, Tian SY, et al. Effects of irrigation management during the rice growing season on soil organic carbon pools. Plant Soil. 2017; 421(1):337–351. https://doi.org/10.1007/s11104-017-3467-7

24. Luo YQ, Zhou XH. Soil Respiration and the Environment. Academic Press. 2006. https://doi.org/10.1016/B978-0-12-088782-8.X5000-1

25. Raubuch M, Behr K, Roose K, Joergensen RG. Specific respiration rates adenylates, and energy budgets of soil microorganisms after addition of transgenic Bt-maize straw. Pedobiologia. 2010; 53:191–196. https://doi.org/10.1016/j.pedobi.2009.10.001

26. Bada D, Marti C, Aguirre AJ. Straw management effects on CO2 efflux and C storage in different Mediterranean agricultural soils. Sci Total Environ. 2013; 465:233–239. https://doi.org/10.1016/j.scitotenv.2013.04.006 PMID: 23642570

27. Zhu LX, Xiao Q, Shen YF, Li SQ. Effects of biochar and maize straw on the short-term carbon and nitrogen dynamics in a cultivated silty loam in China. Environ Sci Pollut Res. 2017; 24(1):1019–1029. https://doi.org/10.1007/s11356-016-7829-0 PMID: 27766524

28. Lenka NK, Lal R. Soil aggregation and greenhouse gas flux after 15 years of wheat straw and fertilizer management in a no-till system. Soil Tillage Res. 2013; 126:78–89. https://doi.org/10.1016/j.still.2012.08.011

29. Peng SL, Guo T, Liu GC. The effects of arbuscular mycorrhizal hyphal networks on soil aggregations of purple soil in southwest China. Soil Biol. Biochem. 2013; 57:411–417. https://doi.org/10.1016/j.soilbio.2012.10.026

30. Zhang J, Hu KL, Li KJ, Zheng CL, Li BG. Simulating the effects of long-term discontinuous and continuous fertilization with straw return on crop yields and soil organic carbon dynamics using the DNDC model. Soil Tillage Res. 2017; 165:302–314. https://doi.org/10.1016/j.still.2016.09.004

31. Jiang X, Boyd CE. Relationship between organic carbon concentration and potential pond bottom soil respiration. Aquacult Eng. 2006; 35(2):147–151. https://doi.org/10.1016/j.aquaculture.2005.10.002

32. Zhao YL, Xue ZW, Guo HB, Mu XY, Li HC. Effects of tillage and crop residue management on soil respiration and its mechanism. Trans Chin Soc Agric Eng. 2014; 30(19):155–165.

33. Pan JL, Dai WA, Shang ZH, Guo RY. Review of research progress on the influence and mechanism of field straw residue incorporation on soil organic matter and nitrogen availability. Chin J Eco-Agric. 2013; 21:526–535.
34. Said-Pullicino D, Cucu MA, Sodano M, Birk JJ, Glaser B, Celi L. Nitrogen immobilization in paddy soils as affected by redox conditions and rice straw incorporation. Geoderma. 2014; 228:44–53. https://doi.org/10.1016/j.geoderma.2013.06.020

35. Zhao X, Wang SQ, Xing G. Nitrification, acidification, and nitrogen leaching from subtropical cropland soils as affected by rice straw-based biochar: laboratory incubation and column leaching studies. J Soils Sediments. 2014; 14:471–482. https://doi.org/10.1007/s11368-013-0803-2

36. Yang HS, Feng JX, Zhai SL, Dai YJ, Xu MM, Wu JS, et al. (2016) Long-term ditch-buried straw return alters soil water potential, temperature, and microbial communities in a rice-wheat rotation system. Soil Tillage Res. 2016; 163:21–31. https://doi.org/10.1016/j.still.2016.05.003

37. Pei PG, Zhang JH, Zhu LF, Hu ZH, Jin QY. Effects of straw returning coupled with N application on rice photosynthetic characteristics, nitrogen uptake and grain yield formation. Chin. J Rice Sci. 2015; 29(3):282–290.

38. Thippayarugs S, Toomsan B, Vityakon P, Cadisch G. Interactions in decomposition and N mineralization between tropical legume residue components. Agroforest Syst. 2008; 72(2):137–148. https://doi.org/10.1007/s10457-007-9062-9

39. Li T, He CE, Ge XY. Ouyang Z. Responses of soil mineral N contents, enzyme activities and crop yield to different C/N ratio mediated by straw retention and N fertilization. Chin J Eco-Agric. 2016; 24(12):1633–1642.

40. Liu J, Huang JH, Yang ZL, Wei F, Liu JG. Soil Respiration variation characteristics of continuous cropping and straw incorporation cotton field. Ecol Environ Sci. 2015; 24(5):791–796.

41. Mao Z. Water-saving irrigation in paddy field plays a key role in production increase and pollution control. China Water Resour. 2009; 21:11–12.

42. Carrijo DR, Lundy ME, Linquist BA. Rice yields and water use under alternate wetting and drying irrigation: A meta-analysis. Field Crop Res. 2017; 203:173–180. https://doi.org/10.1016/j.fcr.2016.12.002

43. Xia LL, Wang SW, Yan XY. Effects of long-term straw incorporation on the net global warming potential and the net economic benefit in a rice–wheat cropping system in China. Agr. Ecosyst Environ. 2014; 197:118–127. https://doi.org/10.1016/j.agee.2014.08.001

44. Hadi A, Inubushi K, Yagi K. Effect of water management on greenhouse gas emissions and microbial properties of paddy soils in Japan and Indonesia. Paddy. Water Environ. 2010; 8(4):319–324. https://doi.org/10.1007/s10333-010-0210-x

45. Hou HJ, Peng SZ, Xu JZ, Yang SH, Mao Z. Seasonal variations of CH4 and N2O emissions in response to water management of paddy fields located in Southeast China. Chemosphere. 2012; 89(7):884–892. https://doi.org/10.1016/j.chemosphere.2012.04.066 PMID: 22673400

46. Li H, Tang QY. Research progress of nitrogen fertilizer use efficiency in paddy fields of China. Crop Res. 2006; (6):401–408.

47. Zhu ZL, Zhang FS. Basic research on nitrogen behaviors and high nitrogen fertilizer use efficiency in Chinese main intensive agricultural ecosystems. Beijing: Science Press, pp:1–27. 2010.

48. Ju XT, Xing GX, Chen XP, Zhang SL, Zhang LJ, Liu XJ, et al. Reducing environmental risk by improving N management in intensive Chinese agricultural systems. Proc Natl Acad Sci U. S. A. 2009; 106:3041–3046. https://doi.org/10.1073/pnas.0813417106 PMID: 19223587