Phylogenetic placement of environmental sequences using taxonomically reliable databases helps to rigorously assess dinophyte biodiversity in Bavarian lakes (Germany)

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Abstract
1. Reliable determination of organisms is a prerequisite to explore their spatial and temporal occurrence and to study their evolution, ecology, and dispersal. In Europe, Bavaria (Germany) provides an excellent study system for research on the origin and diversification of freshwater organisms including dinophytes, due to the presence of extensive lake districts and ice age river valleys. Bavarian freshwater environments are ecologically diverse and range from deep nutrient-poor mountain lakes to shallow nutrient-rich lakes and ponds.
2. We obtained amplicon sequence data (V4 region of small subunit-rRNA, c. 410 bp long) from environmental samples collected at 11 sites in Upper Bavaria. We found 186 operational taxonomic units (OTUs) associated with Dinophyceae that were further classified by means of a phylogenetic placement approach.
3. The maximum likelihood tree inferred from a well-curated reference alignment comprised a systematically representative set of 251 dinophytes, covering the currently known molecular diversity and OTUs linked to type material if possible. Environmental OTUs were scattered across the reference tree, but accumulated mostly in freshwater lineages, with 79% of OTUs placed in either Apocalathium, Ceratium, or Peridinium, the most frequently encountered taxa in Bavaria based on morphology.
4. Twenty-one Bavarian OTUs showed identical sequences to already known and vouchered accessions, two of which are linked to type material, namely Palatinus apiculatus and Theleodinium calcisporum. Particularly within Peridiniaceae, delimitation of Peridinium species was based on the intraspecific sequence variation.
5. Our approach indicates that high-throughput sequencing of environmental samples is effective for reliable determination of dinophyte species in Bavarian lakes. We further discuss the importance of well-curated reference databases that remain to be developed in the future.

KEYWORDS
amplicon sequencing, dinoflagellates, distribution, molecular phylogenetics, operational taxonomic unit
Solid knowledge of ecosystem functioning and community dynamics during seasonal or longer periods, as well as conservation strategies and the impact of invasive species, essentially relies on precise original data about the spatial and temporal occurrence of the inhabiting organisms. Recent advances in the application of high-throughput sequencing (HTS) to analyse the molecular diversity in aquatic environments have enabled a better understanding of the community composition and species distribution in such ecosystems. Within the last decade, our knowledge of the microbial biodiversity has increased faster than ever, especially after Edwards et al. (2006) published the first metagenome analysis of environmental samples using next-generation sequencing technologies. An unexpectedly high protist diversity has been described since then in various aquatic environments (Cuvelier et al., 2010; Keeling et al., 2014; Kohli, Neilan, Brown, Hoppenrath, & Murray, 2014; Massana et al., 2015; Stoek et al., 2010), revealing also the existence of seasonal variation in many phytoplankton taxa (Tillmann, Salas, Jaffrais, Hess, & Silke, 2014; Toebe, Joshi, et al., 2013) and demonstrating the power of these methods in discovering the world’s hidden microbial diversity (Lindeque, Parry, Harmer, Somerfield, & Atkinson, 2013; Medinger et al., 2010). Unknown sequences derived from environmental samples may additionally assist discovering new species and even new lineages (Seevisasan, Sausen, Medlin, & Melkonian, 2013).

One of the most basic and important questions in evolutionary biology of microorganisms refers to mechanisms that have shaped their current distribution (O’Dwyer, Kembel, & Sharpton, 2015). The relative importance of principal processes such as divergence, dispersal, and selection by ecological filtering (Vellend, 2010) is a key point that needs to be worked out rigorously. There is an ongoing debate (Bass & Boenigk, 2011; Caron, 2009; Foissner, 2011) into whether microbes are all cosmopolitan, and lack distinct distributions (Fenchel & Finlay, 2004; Finlay, 2002; Read et al., 2013), or rather follow a moderate endemism model (Bass, Richards, Matthis, Marsh, & Cavalier-Smith, 2007; Bates et al., 2013; Coleman, 2001). Under a less dogmatic perspective, plankton communities may actually consist of both widespread species and those with a smaller range (Bik et al., 2012; Coleman, 2001; Foissner, 2008; Žerdoner Čalasan, Kretschmann, Filipowicz, et al., 2019), and occurrences for such species remain to be identified individually.

A limitation that still precludes a comprehensive knowledge of protist distribution is the taxonomic confusion that exists due to complex determination procedures (Wheeler, 2008). To take full advantage of environmental sequencing, curated reference collections are necessary, linking the molecular data with scientific binominals. Much progress has been made in the past years to build such DNA-barcoding databases, not only for animals and plants (Hollingsworth, 2011; Vernooy et al., 2010), but also for various microorganisms (Quast et al., 2013) such as fungi, diatoms, and dinoflagellates (Del Campo et al., 2018; Fourtanier & Kociolek, 2009; Mordret et al., 2018; Persoh, 2015; Schoch et al., 2012; Stern et al., 2012). These databases are continuously updated and curated (Rigden, Fernandez-Suarez, & Galperin, 2016), preferably based on information from type material (Pawlowski et al., 2012). However, the availability of reference data for environmental sequencing studies is still highly unbalanced across various taxonomic protist lineages.

A prerequisite to any assessment of the microbial biodiversity found in a given ecosystem is the confident placement of resulting molecular operational taxonomic units (OTUs) within particular lineages in the Tree of Life. Due to a huge number of reads, and the diversity of organisms captured by HTS of environmental DNA, initial OTU annotation using reference databases usually allows for confident taxonomic classifications only up to coarser taxonomic ranks (such as phylum, class, or order: Quast et al., 2013). A complementary approach, which enables finer levels of taxonomic assignments (i.e. family, genus, or species), is the phylogenetic placement of reads in reference trees, which also offers the advantage of estimating statistical support values for annotations, regardless of the read’s length (Dunthorn et al., 2014). This method has been recently used in barcoding and biomonitoring projects of protists (Elferink et al., 2017; Keck, Vasselon, Rimet, Bouchez, & Kahler, 2018; Medinger et al., 2010; Vergin et al., 2013).

In comparison to those living in marine environments, the freshwater dinoflagellates are relatively well understood. Currently, 350 species are listed based on morphology (Mertens, Rengefors, Moestrup, & Elleegaard, 2012; Moestrup & Calado, 2018), and continuous efforts to revise them taxonomically have been made in the past years (Moestrup, Lindberg, & Daugbjerg, 2009; Craveiro, Pandeirada, Daugbjerg, Moestrup, & Calado, 2013; Takano, Yamaguchi, Inouye, Moestrup, & Horiguchi, 2014; Kretschmann, Filipowicz, Owsianny, Zinßmeister, & Gottschling, 2015; Kretschmann, Owsianny, Žerdoner Čalasan, & Gottschling, 2018, to mention only a few studies). This also includes the generation of reference DNA sequence information being available for a considerable portion of the species. Nevertheless, there is still a significant portion of freshwater dinoflagellate flora that is neither taxonomically revised nor molecularly investigated. As such, it hinders the transparent research in this field and calls all newly obtained data into question. Moreover, the phenomenon of cryptic speciation (i.e. molecular diversification without morphological differentiation), which is taxonomically challenging in a number of marine dinoflagellates (Hariganeya et al., 2013; Söhner, Zinßmeister, Kirsch, & Gottschling, 2012; Toebe, Alpermann, et al., 2013), appears rare (or understudied) in the freshwater habitat. From the phylogenetic perspective, freshwater dinoflagellates are a heterogeneous group and have colonised their habitats from the marine environment several times independently (Logares et al., 2007; Žerdoner Čalasan, Kretschmann, & Gottschling, 2019). Dinoflagellate development is complex and frequently includes at least two ecologically differentiated stages of life-history, namely a motile theca (vegetative cell) and an immotile, coccosoidal cell (colliquially termed cyst: von Stosch, 1973; Pfiefer & Anderson, 1987).

In Germany, a number of dinoflagellate species have been originally described from the Bavarian region (Baumeister, 1957; Lindemann, 1920; Schrank, 1802) also being the focus of the present study.
However, most algae inhabiting Bavarian waters are presumably wider-ranging, with type localities outside of that region. Mauch, Schmedtje, Maetze, and Fischer (2003) reports 62 dinophyte species (online table available at www.gewaesser-bewertung.de/files/taxaliste.pdf), but this checklist does not provide any source references, is taxonomically not rigorous and also includes species from the marine environment (e.g. Heterocapsa rotundata, which is unlikely to be present in Bavarian freshwater habitats). Morphology based records of Bavarian dinophytes are otherwise rather sporadic, and only a few species have been reported so far including common Ceratium hirundinella, Gyrodinium (≡ Gymnodinium) helveticum, and Peridinium willei (Raeader, 1990; Schaumburg, 1996; Siebeck, 1982), as well as rarer species such as Cystodinium corni f ax and Gloeodinium montanum (Höll, 1928). The true number of dinophyte species in Bavaria, and their spatial occurrences, is thus unknown at present, and more thorough biodiversity assessments of these key protists will benefit from the application of contemporary, high-throughput molecular methods.

In the present study, we provide an initial attempt to apply environmental amplicon sequences with the principal aim to uncover the yet unexplored dinophyte diversity in Bavarian lakes as an exemplary system. We determine their species by placing the ribotypes in phylogenetic reference trees (i.e. tRNA sequence comparison of multiple sequences including GenBank vouchers collected all around the world). Bavarian freshwater environments range from deep nutrient-poor mountain lakes to shallow nutrient-rich lakes and ponds (basic information is available at www.lfu.bayern.de/). These are characterised by a broad range of environmental conditions and resource levels and represent potential habitats for phytoplankton communities. We expect the dinophyte communities in those lakes to be composed of species with different ecological requirements and potentials for dispersal (Zer doner Čalasan, Kretschmann, Filipowicz, et al., 2019), resulting in different effects on their distribution. Our results will lay the basis for a better knowledge of ecosystem functioning and evolutionary dynamics of protists such as freshwater dinophytes.

2 | METHODS

Surface plankton tow samples were collected from piers at 13 localities in Upper Bavaria (Germany) in April 2017 using a plankton net (mesh size 20 μm). The localities included 10 lakes (two lakes were sampled at two sites) and one subsidiary river, to cover standing and flowing bodies of water as well (Table 1, Figure 1). Geographic coordinates were recorded for all sites using a standard GPS Garmin Ltd device. Cells were observed, documented, and measured under a CKX41 inverted microscope (Olympus; Hamburg, Germany) equipped with a phase-contrast option and a DP73 digital camera (Olympus).

Environmental DNA was extracted using the Genomic DNA from Soil kit (Machery-Nagel; Düren, Germany) following the manufacturer's protocol. The small subunit (SSU or 18S) of the ribosomal RNA (rRNA) operon V4 region (c. 410 bp) was the amplification target. Due to polymerase chain reaction (PCR) biases or PCR errors that may artificially increase diversity, each PCR was performed in triplicates (Aca nis, Sarma-Rupavattm, Klepac-Ceraj, & Polz, 2005). Forward and reverse primers were those used by Xiao, Wu, Liu, Xu, and Chi (2017). Amplification of DNA (PCR) for subsequent amplicon sequencing (Illumina) was carried out using 5 ng/μl template DNA, 1 μM of each primer and 2× KAPA Hifl HotStart Ready Mix (Roche; Penzberg, Germany). Resulting PCR products were visualised in 1% agarose gels and purified using AMPure XP Beads (Beckman Coulter). Dual indices and Illumina sequence adapters were attached by means of an Index PCR using the Nextera XT Index Kit (Illumina), and final PCR products were again purified using AMPure XP Beads. The library was validated using an Agilent 2100 Bioanalyzer Software and a DNA 1000 Chip (Agilent Technologies) to verify the size of the resulting fragments. The final DNA libraries were equimolarly pooled and run in a MiSeq System (Illumina) after combining the denatured PhiX control library (15%) and the denatured amplicon library. Some 6.5 million 2 × 300 bp paired-end reads were produced and demultiplexed into 13 samples.

Using Trim momatic (v0.38; Bolger, Lohse, & Usadel, 2014), 3'-ends of the reads were trimmed based on read quality information. PEAR (v0.9.10; Zhang, Kobert, Flouri, & Stamatakis, 2014) with default settings was used to merge the paired-end reads. Sequences that could not be merged were discarded. Primer-matching sequence segments were truncated from the amplicons by cutadapt (v1.9; Martin, 2011) and amplicons were only kept in the sequence pool if both the segments of the forward and of the reverse primer could be found. Remaining sequences were filtered for further quality features by vsearch (v2.3.0; Rognes, Flouri, Nichols, Quince, & Mahé, 2016). Sequences were discarded if they were outside a 50 bp radius above or below the median length of the primer-truncated amplicon (c. 387 bp), if they carried any ambiguity, or if the expected number of miscalled bases of a sequence (sum of all base error probabilities of a sequence) was >1. Chimera were predicted also by vsearch utilising the UCHIME algorithm (Edgar, Haas, Clemente, Quince, & Knight, 2011) with default settings in de novo mode for each sample separately and removed from the sample files. About 4 million sequences passed all filtering steps and were used as input for the OTU-clustering, which was done using the tool Swarm (v2.1.8; Mahé, Rognes, Quince, De Vargas, & Dunthorn, 2015) with default settings. The most abundant amplicon of each OTU-cluster was used as an OTU representative. These sequences were annotated by the RDP classifier (Wang, Garrity, Tiedje, & Cole, 2007) implemented in mothur (v1.38.1; Schloss et al., 2009) using the Ref_NR99 version of release 128 of the SILVA SSU sequence set (Quast et al., 2013) using a reference with a confidence cut-off of 90. The annotation of each representative sequence was used as annotation of the OTU-cluster as well and added to the corresponding line of the OTU table.

Operational taxonomic units classified as Dinoflagentata as search strings and with an abundance of ≥10 were classified more accurately by phylogenetic placement onto a reference tree based on concatenated rRNA alignments. Full voucher information of this
**TABLE 1** Geographic origin of the samples used in the molecular analyses. To characterise the localities in terms of their ecology, we provide the corresponding trophic state index (Nürnberg, 1996; OECD, 1982)

| Collectors | Collection number | Locality (trophic state index) | Geographic coordinates (latitude, longitude) | Area (km²) | Max. depth (m) | Mean depth (m) | Altitude (m) |
|------------|------------------|--------------------------------|----------------------------------------------|-----------|---------------|---------------|-------------|
| J. Kretschmann & J. Schmid | D075 | Lake Kochel (mesotrophic peri-alpine lake) | 47°39.200’N 11°19.001’E | 3.95 | 66 | 31 | 601 |
| J. Kretschmann & J. Schmid | D076 | Riegsee (mesotrophic peri-alpine lake) | 47°41.395’N 11°13.849’E | 1.97 | 15.4 | 6.8 | 661 |
| J. Kretschmann & J. Schmid | D077 | Staffelsee (oligotrophic peri-alpine lake) | 47°41.406’N 11°10.631’E | 7.66 | 39.4 | 9.8 | 648 |
| A. Žerdoner Čalasan & M. Gottschling | D078 | Starnberger See (mesotrophic peri-alpine lake) | 47°41.032’N 11°11.342’E | 56.36 | 127.8 | 53.2 | 533 |
| A. Žerdoner Čalasan & M. Gottschling | D079 | Eitzenberger Weiher (eutrophic moor pond) | 47°47.122’N 11°21.757’E | 0.0534 | 10 | Unknown | 602 |
| A. Žerdoner Čalasan & M. Gottschling | D080 | Oster See (mesotrophic peri-alpine lake) | 47°47.687’N 11°18.357’E | 0.235 | 29.7 | 9.38 | 589 |
| A. Žerdoner Čalasan & M. Gottschling | D081 | Hardtbach near Haarsee (mesotrophic water reservoir) | 47°49.044’N 11°13.206’E | 0.0647 | 8.8 | Unknown | 643 |
| A. Žerdoner Čalasan & M. Gottschling | D082 | Zelsee (mesotrophic water reservoir) | 47°52.335’N 11°03.848’E | 1 | <5 | Unknown | 583 |
| A. Žerdoner Čalasan & M. Gottschling | D083 | Lech at Apfeldorf (peri-alpine river bay) | 47°54.120’N 10°55.900’E | – | – | – | 640 |
| A. Žerdoner Čalasan & M. Gottschling | D084 | Ammersee (Diessen) (mesotrophic peri-alpine lake) | 47°57.017’N 11°06.571’E | 46.6 | 81.1 | 37.5 | 530 |
| A. Žerdoner Čalasan & M. Gottschling | D085 | Ammersee (Herrsching) (mesotrophic peri-alpine lake) | 47°59.570’N 11°10.188’E | 46.6 | 81.1 | 37.5 | 531 |
| A. Žerdoner Čalasan & M. Gottschling | D086 | Pilsensee (mesotrophic peri-alpine lake) | 48°01.421’N 11°11.764’E | 1.95 | 17.1 | 9.3 | 525 |
| A. Žerdoner Čalasan & M. Gottschling | D087 | Lake Wörth (mesotrophic peri-alpine lake) | 48°04.110’N 11°12.013’E | 4.34 | 34 | 14.7 | 553 |
systematically representative set comprising 241 dinophytes (plus 10 outgroup accessions) are provided in Table S1. To further explore dinoflagellate identities based on DNA sequences, we performed BLAST searches (Altschul, Gish, Miller, Myers, & Lipman, 1990.

For alignment constitution, separate matrices of the rRNA operon (i.e. SSU, ITS, LSU) were constructed, aligned using "MAFFT" v6.502a (Katoh & Standley, 2013) and concatenated afterwards. The aligned matrices are available as *.nex files upon request. Dinoflagellate phylogenetic analyses were carried out using maximum likelihood (ML) and Bayesian approaches, as described in detail previously (Gottschling et al., 2012) using the resources available from the CIPRES Science Gateway (Miller, Pfeiffer, & Schwartz, 2010). The Bayesian analysis was performed using "MrBayes" v3.2.6 (Ronquist et al., 2012, freely available at http://mrbayes.sourceforge.net/download.php) under the GTR+Γ substitution model and the random-addition-sequence method with 10 replicates. We ran two independent analyses of four chains (one cold and three heated) with 20,000,000 generations, sampled every 1,000th cycle, with an appropriate burn-in (10%) as inferred from the evaluation of the trace files using Tracer v1.5 (http://tree.bio.ed.ac.uk/software/tracer/). For the ML calculation, the MPI version of "RAxML" v8.2.4 (Stamatakis, 2014, freely available at http://www.exelixis-lab.org/) was applied using the GTR+Γ substitution model under the CAT approximation. We determined the best-scoring ML tree and

FIGURE 1 Selected Bavarian lakes under investigation. Collection sites are indicated by red dots, GPS coordinates are given in Table 1. The map was made using geographic data for the inland waters in Germany, which are freely available at the DIVA-GIS website (http://www.diva-gis.org/gdata, accessed 12 April, 2019) and the free software QGIS, Free Open Source Geographic Information System (QGIS Development Team, 2019).
performed 1,000 non-parametric bootstrap replicates (rapid analysis) in a single step. Statistical support values (LBS: ML bootstrap support, BPP: Bayesian posterior probabilities) were drawn on the resulting, best-scoring tree. To assess whether certain amplicons are nested in other freshwater clades, or belong to yet unidentified lineages deriving from marine clades, we focussed on phylogenetic context of reads with >95% similarity to known species.

3 | RESULTS

In the plankton tow samples analysed using light microscopy, we observed a considerable morphological diversity of dinophytes (Figures 2 and 3), comprising about 10 different species and 2–3 easily distinguishable species per locality. It included photoautotroph taxa such as the relatively easy recognisable Apocalathium (Figure 2f,g), Ceratium (Figure 3), and Peridinium (Figure 2d,h–n), but also heterotroph taxa such as a Pfiesteria-like dinophytes (Figure 2e). Occasionally, coccoid developmental stages of dinophytes were also observed (Figure 2b–d).

The rRNA reference alignment of dinophytes was 1,857 + 1,670 + 3,745 bp long and was composed of 965 + 839 + 1,963 parsimony-informative sites (51.8%, mean of 15.0 per terminal taxon) and 5,308 distinct RAxML alignment patterns. Figure 4 shows the best-scoring ML tree (−ln = 232,502.267554), with the internal topology not fully resolved. However, Dinophyceae were monophyletic (99LBS), many nodes were statistically well if not maximally supported, and a number of lineages at high taxonomic
level such as Dinophysales (100LBS, 1.00BPP), Gonyaulacales, Gymnodiniales (75LBS, 1.00BPP), Peridiniales, Prorocentrales and Suessiales (91LBS, 1.00BPP), but also Amphidinomataceae (95LBS, 1.00BPP), Brachidiniaceae (97LBS), and Tovelliaceae (82LBS, 1.00BPP) were recognised. Only 13 of 241 dinofytes (5.4%) were not assigned to any of such lineages. At least 21 distantly related lineages of freshwater dinofytes were scattered over the reference tree, showing a polyphyletic pattern.

In total, we obtained 1,155 OTUs classified as dinofytes, most of which were singletons, doubletons, or occurred in frequencies of <10 (data available upon request). All subsequent work was performed with those 186 dinofyte OTUs with an abundance of ≥10. The vast majority of environmental OTUs were assigned to the photoautotrophic dinofytes and were placed mostly in the freshwater lineages (Figure 5). Large amounts of OTUs were determined as Apocalathium (16.1%), Ceratium (45.2%), or Peridinium (17.7%). However, 18 OTUs (9.7%) were assigned to so far marine taxa including, for example, Blastodinium, *†Calcicaripinum* (both Peridiniales), *Dinophysys* (Dinophysales), and Tropicos (Gonyaulacales). A number of OTUs exhibited considerable phylogenetic distance to known vouchered sequences. These were found among Blastodiniaceae, Borghiellaceae, Gonyaulacaceae and particularly Pfiesteria-like dinofytes (Thoracosphaeraceae; not shown).

Figure 6 shows a phylogeny, as a cut-off from the reference tree plus the environmental OTUs, including *Peridinium* with considerable sequence variation. All respective OTUs were placed in one of the established species *Peridinium bipes*, *Peridinium cinctum*, *Peridinium gutunense*, or *Peridinium wiliei*, each based on statistical support values ≥ 75LBS and .92BPP, respectively. Twenty-one OTUs (11.3%) were (almost) identical to known (reference) sequences, and were therefore reliably assigned to, for example, *Naiadinium polonicum* and *Theleodinium calcisporum* from the Thoracosphaeraceae as well as *Palatinus apiculatus* from the Peridiniosidae (Figure 6) and *Biecheleria brevisulcata* from the Suessiaceae (the latter three species are linked to type material: Table S1). Five OTUs identical to vouchered sequences were assigned to the marine species “*Alexandrium hiranoi*, *Lingulodinium polyedra* (both Gonyaulacales), *Phalacroma rotundatum* (Dinophysales), *Pelagodinium bei*, and *Polarella glacialis* (both Suessiales). Reliable species determinations of our study are summarised in Table 2.

### 4 | DISCUSSION

Reliable determination of organisms is a necessary prerequisite to explore their spatial and temporal occurrence and to rigorously test hypotheses on their diversification, ecology, and dispersal. Flowering plants, insects, and larger animals are well represented in extensive collections (Krupnick & Kress, 2005; Mayer et al., 2013; Rocha et al., 2014; Steinicke, 2014). In numerous cases, these have also been digitised over the course of the past decade, providing enduring and exact publicly available occurrence data (e.g. GBIF, GBOL, JSTOR, Tropicos®). Such powerful and continuously curated databases are scarce for protists, which are too small for direct observation and need microscopic expertise for examination. However, the problem is recognised, and considerable efforts have been made to build curated sequence databases and reference phylogenetic trees for dinofytes (Del Campo et al., 2018; Elferink et al., 2017; Mordret et al., 2018; Quast et al., 2013) and other microbial taxa.

Our reference tree comprising the known dinofycean sequence diversity is largely in agreement with previous rRNA approaches (Gu et al., 2013) as well as those based on excessive transcriptome sequence data (although using a much smaller taxon sample: Janouškovec et al., 2017; Price & Bhattacharya, 2017). The tree reflects traditional taxonomic concepts based on morphology better than previous (frequently single-locus) approaches, and we are able to recognise major monophyletic lineages of dinofytes. Trees such as that obtained in this study serve as a taxonomic scaffold for the systematic placement of newly generated dinofyte sequences, even if they are short as it is in the case of next-generation
sequencing amplicons. Our approach to determine dinophyte species using reference trees as inferred from multi-locus rRNA alignments is proven successful to a certain degree, at least for samples from the freshwater environment (providing also some new dinophyte records for Bavaria: Table 2), but also from the marine realm (Elferink et al., 2017; Wohlrab et al., 2018). Only if taxonomically curated and kept up to date can the rRNA alignments also serve as a tool for placing other unknown protist environmental sequences into the Tree of Life. However, such rRNA-based reference databases are still in their infancy for most other protist lineages.

**FIGURE 4** A molecular reference tree recognising major groups of dinophytes. Maximum likelihood (ML) tree of 241 systematically representative dinophyte sequences (with strain number information) as inferred from a rRNA nucleotide alignment (3,767 parsimony-informative positions). Numbers on branches are ML bootstrap (above) and Bayesian support values (below) for the clusters (asterisks indicate maximal support values, values under 50 and 0.90, respectively, are not shown). Colour coding: black, marine OTU; red, freshwater OTU.
The most frequently encountered taxa correspond to those that are known from morphological surveys in Bavaria, which comprise *Apocalathium* (Mauch et al., 2003; Mischke, Riedmüller, Hoehn, Denek, & Nixdorf, 2015; Schaumburg, 1996), *Ceratium* (Mauch et al., 2003; Raeder, 1990; Schaumburg, 1996; Schaumburg & Hehl, 2001), and *Peridinium* (Fröbrich, Mangelsdorf, Schauer, Streil, & Wachter, 1977; Mauch et al., 2003; Mischke et al., 2015; Raeder, 1990; Schaumburg, 1996), summing up to almost 80% of all identified OTUs. Our sequence-based findings seem to support our own morphological confirmations in the samples that we have investigated, and such combinatorial approaches are needed (Medinger et al., 2010; Rimet, Vasselton, A-Keszte, & Bouchez, 2018; Mora et al., 2019), as long as the taxonomic impediment and uncertainty with taxon determination continue to exist. However, when comparing morphological and genetic species diversity assessments, one has to take into account the morphological variation and the cryptic speciation. On a broader spectrum, these might namely either seemingly corroborate or contradict each other. Thus, having a broad knowledge on the biology of investigated microbiota is of great importance, as for the example given below.

The Peridiniaceae are one of the most important groups of freshwater dinophytes and may comprise about a dozen species (Gottschling, Kretschmann, & Žerdoner Čalasan, 2017; Moestrup & Calado, 2018), half of which are already known from molecular DNA sequences (Table S1). In rRNA sequences, they show a considerable variation even within species (Izquierdo López, Kretschmann, Žerdoner Čalasan, & Gottschling, 2018) leading to long branches in phylogenetic trees (Gu et al., 2013). This intraspecific variability, in combination with a reliable taxonomy at the species level, makes the Peridiniaceae a good example of how species determination can be effective using environmental amplicon sequences. We are able to assign all peridiniacean OTUs gained in this study to an established species of *Peridinium* (at least as long as no cryptic speciation has been documented in this lineage). *Peridinium bipes* exhibits a distinct morphology, but our approach allows for efficient differentiation between, for example, *P. cinctum* and *P. willei* that are challenging to tell apart using just light microscopy in monitoring studies. However, sequence variation might be lower (and closely related species may exhibit identical SSU sequences) in other groups such as *Scripsiella*, and it is therefore not always possible to differentiate between

**FIGURE 5** Variation in dinophyte abundances in Bavarian lakes. Abundance of environmental OTUs in the different collection sites, with names based on phylogenetic placement. GON, Gonyaulacales; GYM, Gymnodiniales; PER, Peridiniales; SUE, Suessiales
species based on one particular molecular locus (Söhner et al., 2012; Žerdoner Čalasan, Kretschmann, Filipowicz, et al., 2019). Because of the large heterogeneity of evolutionary rates in the rRNA operon (Gu et al., 2013; Saldarriaga Echavarría, Taylor, Cavalier-Smith, Menden-Deuerd, & Keeling, 2004; Stern et al., 2012; Žerdoner Čalasan, Kretschmann, & Gottschling, 2019), there is no such thing as an universal DNA barcode for all constituting lineages, and other loci including ITS or LSU may prove better suited for reliable species determination of such organism sets using HTS.

The potential identification of few marine taxa in the freshwater environment based on sequences comes as a surprise. Contamination due to the applied methods is very unlikely, as these species are not cultivated in our laboratories, but the incomplete assessment of freshwater dinophytes in DNA reference databases may account for the ostensible detection of marine taxa in our samples. However, some dinophytes have colonised freshwater from marine habitats relatively recently (Žerdoner Čalasan, Kretschmann, & Gottschling, 2019). Furthermore, different rates of the rRNA operon evolution...
TABLE 2  Occurrence of selected dinophytes that could be reliably determined to the species level based on environmental rRNA sequence data. DIN, Dinophysales; GON, Gonyaulacales; GYM, Gymnodiniales; PER, Peridiniales; SUE, Suessiales. Taxa not listed in Mauch et al. (2003) are indicated with an asterisk.

| Species (taxonomic assignment)       | Ammersee (Diessen) | Ammersee (Herrsching) | Ettendorfer Weiher | Hardtbach near Haarsee | Lake Kochel | Lech at Apfeldorf | Oster See | Pilisensee | Riegsee | Staffelsee | Starnberger See | Lake Wörth | Zellsee |
|--------------------------------------|--------------------|-----------------------|-------------------|------------------------|-------------|------------------|-----------|------------|---------|-----------|----------------|-------------|---------|
| "Alexandrium" hiranoi*               |                    |                       |                   |                        |             |                  |           |            |         |           |                 |             | +       |
| (GON)                                |                    |                       |                   |                        |             |                  |           |            |         |           |                 |             |         |
| Asulcocephalium miricentronis* (SUE) |                    |                       |                   |                        |             |                  |           |            |         |           |                 |             | +       |
| Biecheleria brevisulcata (SUE)       | +                  |                       |                   |                        |             |                  |           |            |         |           |                 |             |         |
| Chimonodinium lomnickii (PER)        | +                  |                       |                   |                        |             |                  |           |            |         |           |                 |             |         |
| Lingulodinium polyedra*              |                    |                       |                   |                        |             |                  |           |            | +       |           |                 |             | +       |
| (GON)                                |                    |                       |                   |                        |             |                  |           |            |         |           |                 |             |         |
| Naiadinium polonicum (PER)           | +                  |                       |                   |                        |             |                  |           |            | +       |           |                 |             |         |
| Polatinus apiculatus (PER)           |                    |                       |                   |                        |             |                  |           |            | +       |           |                 |             |         |
| Peridinium bipes (PER)               | +                  |                       |                   |                        |             |                  |           |            | +       |           |                 |             |         |
| Peridinium cinctum (PER)             | +                  |                       |                   |                        |             |                  |           |            | +       |           |                 |             |         |
| Peridinium gatunense (PER)           | +                  | +                     |                   |                        |             |                  |           |            | +       |           |                 |             | +       |
| Peridinium willei (PER)              | +                  | +                     |                   |                        |             |                  |           |            | +       |           |                 |             | +       |
| Phalacroma rotundatum*               |                    |                       |                   |                        |             |                  |           |            | +       |           |                 |             |         |
| (DIN)                                |                    |                       |                   |                        |             |                  |           |            | +       |           |                 |             |         |
| Polarella glacialis* (SUE)           |                    |                       |                   |                        |             |                  |           |            | +       |           |                 |             | +       |
| Spiniferodinium palustre (GYM)       |                    |                       |                   |                        |             |                  |           |            | +       |           |                 |             |         |
| Theleodinium calcisporum* (PER)      |                    |                       |                   |                        |             |                  |           |            | +       |           |                 |             |         |
| Unruhdinium penardii (PER)           |                    |                       |                   |                        |             |                  |           |            | +       |           |                 |             | +       |
may impact the phylogenetic position of certain reads. For instance, *Gonyaulax clevei* has been reported from German lakes (Hickel & Pollhinger, 1986), but has never been sequenced. It is possible that those sequences once gained will group with marine taxa such as species of *Lingulodinium*. Moreover, a few organisms are physiologically able to successfully overcome the physiological barrier between the oceans and freshwater habitats (Pokorný, 2009) and establish new populations. As living models, species such as *Huia caspica* and *Kolkwitzia acuta* might be a key in this respect, as they are found in both marine and freshwater habitats (Gu, Mertens, & Liu, 2016; Mertens et al., 2015). By any means, the precise biological role, and the overall biological activity, has to be worked out for marine taxa in freshwater environments. Future research should rigorously use physical specimens and preferably living strains. Moreover, dinoflagellate species richness in Bavarian lakes may be greater than that previously reported in the literature based on our genetic analysis. Particularly, the considerable number of so far unknown *Pfiesteria*-like sequences (Burkholder & Marshall, 2012; Calado, Cravero, Daugbjerg, & Moestrup, 2009; Litaker et al., 2005) is impressive. A targeted search in future will assess this diversity and address questions, such as whether these are already accepted species without hitherto known DNA sequence information or even new species.

In the microbial world, the importance of DNA sequences linked to type material cannot be overestimated. In this respect, our approach to place OTUs on a reference tree using curated and voucheded representatives has been proven successful with the documentation of sequences identical to *Biecheleria brevisulcata* (Suessiaceae), *Palatinus apiculatus* (Peridiniopsidaceae), and *Theleodinium calcisporum* (Thracosphaeraceae). The first and latter species have been described only a few years ago (Cravero et al., 2013; Takahashi, Sarai, & Iwataki, 2014), but *Palatinus apiculatus* refers to a historical name from the 19th century (Ehrenberg, 1838). Usually, such names are not linked to DNA sequence information, but the application of epitypification has made the determination of such species unambiguous (Kretschmann, Žerdoner Čalasan, Kusber, & Gottschling, 2018). This strategy has not—to the best of our knowledge—been applied for the species of *Ceratium*, which makes the determination of numerous OTUs gained in this study at the species level vague. Once identified, strategic taxonomic classifications of target organisms are possible and may refer to names such as *Ceratium macroceras* and *Ceratium tetraceros*, both being described from Bavaria (Schrank, 1793, 1802). However, as the potential intraspecific and interspecific variability together with possible cryptic speciation and morphological variability of *Ceratium* are currently unknown, the taxonomic clarification must be performed with great caution, taking all such aspects into account. The taxonomic tool of epitypification provided by the International Code of Nomenclature for algae, fungi, and plants (Turland et al., 2018) is still not regularly used in phycology, despite being of great importance for stable taxonomy. Such approaches must be carried out, because the resolution power of all available HTS methods, as exemplified in this study, always comes down to the basic necessity of a reliable application of names.

Curated contemporary reference databases leave further room for improvements. For example, a number of the OTU sequences obtained in this study were assigned to *Scrippsiella* by the SILVA (Quast et al., 2013) reference, which is a predominantly marine dinoflagellate lineage. Such entries misleadingly refer to it as "Scrippsiella" *hongoei*, which is a name being classified today under *Apocalathium*. Therefore, the correct species name for the OTUs is (probably) freshwater *Apocalathium acicularifurum*. It might be only a matter of time until this particular taxonomic confusion is corrected for future releases of the SILVA databases, as this error has been already corrected in the DI NOREF database (Mordret et al., 2018). The latter, in turn, relies on SSU reference sequence data only and is therefore unable to place environmental OTUs of studies using LSU (Elferink et al., 2017) and/or ITS sequences (Lutz, McCutcheon, McQuaid, & Benning, 2018). Concomitantly, taxa such as *Peridiniella*, *Sphaerodinium* and *Tyrannodinium*, of which only LSU sequences are known at present, are subsequently undetectable using DI NOREF (however, †Leonella and †Posoniella are also missing, although SSU reference sequences are already available). The variety of sequencing approaches thus requires a database that assures both, provision of extensive rRNA sequence information and taxonomic reliability. In this respect, the indication of sequences that have been gained from type material is also important (Pawlowski et al., 2012) and needs to be added in DI NOREF. Moreover, curation of the DI NOREF database has not prevented the use of names such as ‘*Goniodomataceae*’, which have been nomenclaturally rejected (Elbrächter & Gottschling, 2015; Prud‘homme Van Reine, 2017).

Our approach to detect dinoflagellates in Bavarian lakes is powerful and will lay the basis for solid information on which species are widely distributed and abundant, and which species are rarer and represent rather endemic entities with narrower distributions. If occurrences of dinoflagellate species correlate with environmental traits, then improved species circumscriptions also taking their ecological niche into account are possible. With our project, we may start to understand not only that a certain species occurs in a given freshwater habitat, but also *why*. Our example of field mapping, and the pursued predictability of freshwater dinoflagellate occurrences, has thus a great potential to serve as a model for other taxonomic groups and/or the investigation of similar and alternative environments in other parts of the world.

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