Added value recyclability of glass fiber waste as photo-oxidation catalyst for toxic cytostatic micropollutants

Gheorghe Nechifor¹, Eugenia Eftimie Totu¹*, Aurelia Cristina Nechifor², Alina Mirela Constantin², Mihaela Elena Cărăuşu³ & Ibrahim Isildaκ⁴

There is an increased interest in recycling valuable waste materials for usage in procedures with high added values. Silica microparticles are involved in the processes of catalysis, separation, immobilization of complexants, biologically active compounds, and different nanospecies, responding to restrictive requirements for selectivity of various chemical and biochemical processes. This paper presents the surface modification of accessible and dimensionally controlled recycled silica microfiber with titanium dioxide. Strong base species in organic solvents: methoxide, ethoxide, propoxide, and potassium butoxide in corresponding alcohol, activated the glass microfibres with 12–13 µm diameter. In the photo-oxidation process of a toxic micro-pollutant, cyclophosphamide, the new composite material successfully proved photocatalytic effectiveness. The present work fulfills simultaneously two specific objectives related to the efforts directed towards a sustainable environment and circular economy: recycling of optical glass microfibers resulted as waste from the industry, and their usage for the photo-oxidation of highly toxic emerging micro-pollutants.

Inorganic-inorganic oxide-type composite materials offer the most interesting technical solutions in many physicochemical processes of bio-degradation, coating, separation, or catalysis. A particular case is presented by the silicon dioxide-titanium dioxide (TiO₂-SiO₂) composites that combine the distinctive characteristics of the two oxides. Thus, silicon dioxide is a technically-economically accessible material and can be made in the form of spherical particles, nanofibers, nanotubes, or films, thus being used as an active material, adsorbent or support for other materials with higher selectivity. Titanium dioxide prepared in predetermined forms: spheres, nanotubes, nano-threads, or films could be used as an adsorbent or covering, but especially like catalytic material. Of course, the combination of the two oxides' properties has been used for the most advanced applications: photocatalytic degradation of organic/ pharmaceutical pollutants, fuel cells, membrane reactors, bioreactors, advanced separations and clinical devices.

Environmental researches proved that the widely used cytostatic drug, N, N-bis(2-chloroethyl)-1,3,2-oxazaphosphinan-2-amine 2-oxide known as cyclophosphamide, which is a cyclic amide, produces residues with mutagenic action. Although mainly known as an anticancer drug, cyclophosphamide is also administrated frequently as an anti-inflammatory or immunolytic agent for rheumatoid arthritis, or nephrotic syndrome, systemic lupus erythematosus, or systemic sclerosis lung fibrosis. This important cytostatic compound administered for autoimmune diseases, and cancer chemotherapy, as well as its metabolites, are released into effluents that reach surface and ground waters. The cyclophosphamide action through nucleophilic compounds' alkylation and metabolic activation result in dramatic side-effects like carcinogenic behavior and genotoxic effect. Figure 1 presents the specific action and the activation mechanism of cyclophosphamide.

¹Faculty of Applied Chemistry and Material Science, Polytechnic University of Bucharest, 060042, Bucharest, Romania. ²National Research and Development Institute for Industrial Ecology – ECOIND Bucharest, 71-73 Drumul Podul Dambovitei Str., 060652, Bucharest, Romania. ³Department of Public Health and Management, Faculty of Dental Medicine, Grigore T. Popa University of Medicine and Pharmacy, 700115, Iasi, Romania. ⁴Department of Bioengineering, Faculty of Chemical and Metallurgical Engineering, Yildiz Technical University, 34210, Esenler-Istanbul, Turkey. *email: eugenia.totu@upb.ro
The action mechanism of cyclophosphamide, according to Figure 1, can provoke the DNA damage. The alkylating drug bio-action is triggered by P450 cytochrome when results 4-hydroxyl cyclophosphamide. Following a ring-opening reaction, the metabolically activated cyclophosphamide will further generate the aldotrophamide, which forms the phosphoramide mustard (active metabolite) and acrolein25. The cyclophosphamide's metabolite, phosphoramide mustard, presents anticancer activity despite its toxicity. The cytotoxic activity is due to the protein and DNA alkylation generated by the cyclophosphamide reactive metabolites26. The activated metabolites, as shown in Figure 1, act on both healthy and cancerous cells.

In Figure 1, there is signaled the formation of reactive oxygen species (ROS). The action of either some cytochrome or mitochondria – mitochondrial pathway could explain the ROS presence. When the oxidative forces imbalanced the antioxidant system action of the body, oxidative stress occurs. Such a detrimental body state is due to the presence of free radicals generated by the reactive oxygen metabolites. In vivo, oxygen molecules are reduced, forming highly reactive oxygen species as hydrogen peroxide, superoxide radical, or alkoxyl radicals, among others. In the case of excess production of such highly reactive and unstable radicals, the body antioxidant defense mechanism is overwhelmed by the reactive oxygen species presence and the oxidative stress occurs27. The reactive oxygen species' accumulation could initiate cell apoptosis. As a consequence, the usage of cyclophosphamide was proved to generate oxidative stress, finally inducing cell apoptosis26.

Therefore, being a hazardous drug, cyclophosphamide has to be removed from any water, as it considered to be a toxic micro-pollutant due to its adverse effects on human health through the drinking water28 and wastewaters29. This micro-pollutant is mainly present in wastewaters due to the resulting industrial used waters from the organic synthesis of the cytostatic drugs where could be identified the cyclophosphamide's specific active principles. The patients treated with cyclophosphamide release through the biological fluids (urine) either the drug metabolites, either a part of the untransformed cytostatic drug. Therefore, a second important pollution source is represented by the discharged wastewaters from the oncological treatment clinics. The currently applied treatments of the industrial and clinical resulting wastewaters suppose the application of different physical and chemical processes, such as adsorption, chemical and UV oxidation30, membrane separations, flocculation, biodegradation31, which are costly and do not provide a specific efficiency for reducing the cyclophosphamide concentration to the legally imposed levels – less than 1 µg/dm³.

The conventional method applied, the ozonation, inactivates the cyclophosphamide quite slowly, its destruction taking place in time. In consequence, such cytostatic drug and its metabolites could remain in surface or ground waters, and from there, they could be found in drinking water. Experimental data showed that about 10 g cyclophosphamide intake during 3–6 months rise by 1.5 factor the environmental risk, while the threshold level based on a risk assessment data for active carcinogenic compounds is 10 ng/L32. Selective determinations highlighted that through drinking water, there is a possible intake lifetime risk of 10⁻² or less than 10⁻⁶33. However, as mentioned above, not always the cyclophosphamide could be efficiently biodegraded. In such a context, it is of utmost importance to develop affordable technologies for limiting and reducing the presence of cyclophosphamide into wastewaters34.

Although the synthesis of silicon dioxide in various geometric forms has been extensively studied and available technical solutions have been found at industrial scale35, some economic aspects determined in particular by the high energy consumption and the environmental protection remain a significant challenge. The specific case...
of using silicon dioxide as support in titanium dioxide composites raises the problem of obtaining the silicon in the predetermined form under economically favorable conditions and, if possible, also protecting the environment. For reproducible silica geometrically shaped, we could consider as reliable and abundant source the optical glass fibers manufacturing industry. The European, US, and Asia-Pacific markets for glass fiber increased tremendously during the last years. For instance, the forecast for 2023 of the specific European market is 2.1 billion USD, with an annual growth rate increasing by 2.8% compared with the 2018 glass fiber production. Also, the actual trend of the global glass fiber’s market points toward an annual growth rate exceeding 4% for the period 2019 to 2024. Such significant development of the glass fiber industry raises environmental concerns related to the glass fibers’ waste disposal. Recycling supposes to transform the glass fiber waste into a new product suitable for returning into the economic mainstream and being financially viable in order to implement the proposed procedure. For instance, the environmental legislation, as is the European Union Directive 2008/98/EC set high targets for recycling techniques. The efforts to develop new recycling methods are welcome, much more if such an applied process would allow environmental applications instead of the environmental burden represented by the glass fiber waste. Recognizing the global production level of 325 million fiber kilometers for 2018, the level of waste resulting from cutting the fibers to a preset size became important.

Seeking for a clean environment, the balance between the environment’s protection and manufacturing industry, became a strategic aim. Thus, our research circumscribes to sustained efforts to find viable solutions for ecosystem protection.

In this paper, we propose to re-use silica microfibers (quartz) – a valuable waste from the optical glass fibers industrial facilities, as support material for titanium dioxide to obtain a composite material through the sol-gel method: silica microfibers with adherent titanium dioxide deposits. The efficiency of new composite material has been tested in the photo-oxidation process of cyclophosphamide in the effort to develop an appropriate method for degrading the cyclic amides with carcinogenic, genotoxic, or mutagenic potential. A new heterogeneous (SiO2 – TiO2 – UV irradiation) treatment solution is welcome, taking into account the imposed limit under 1 μg/L in discharged wastewaters from the drug synthesis industries or cancer treatment clinics for such anticancer drugs.

**Results**

Complex characterization of quartz fibers decorated with TiO2. Morphological and structural analysis of SiO2 microfibers decorated with nanoTiO2 particles. The quartz microfibers used in this study represent a technical–economically accessible source being a by-product of the silica fiber industry whose use for obtaining advanced materials dedicated to various physicochemical processes: selective separations, controlled adsorption, combustion cells or photo-oxidation reactors would be a niche of valorization. The studied silica microfibers have a morphology (as shown in Supplementary Material – Figure S.1.) constant and ideal for use as a support for active or selective materials as adsorbsents or catalysts. The scanning electron microscopy put in evidence the smooth surface of microfibers with 12–13 μm diameter and 100 μm length. Also, the EDX analysis presented in Supplementary Material – Figure S.2, evidenced for the silica microfibers a controlled composition.

The grafting of titanium dioxide microparticles on silica microfibers was carried out by *in situ* titanium dioxide generation through the sol-gel method in alcohol and using the appropriate titanates: tetraethoxy-, tetraethoxy-, tetrapropoxy- and tetra-butylated titanate.

In Figure 2 are presented the morphological features of the quartz microfibers decorated with titanium dioxide resulting from the tetra-butyl-titanate - tert-butyl alcohol couple. The grafted titania presented dimensions of 1–3.5 μm, while the microfibres diameters are slowly increasing towards 14 μm (Figure 2a).

The EDX spectrum, Figure 2b, highlights the composition of the nanostructured microparticles deposited onto silica microfibers. The diagram puts in evidence the presence of titanium dioxide.

Figure 2c displays the XRD pattern of the titanium dioxide that decorates the silica microfibers. The X-rays diffractogram highlights the presence of a mixed-phase for nanotitania, being recorded the characteristic peaks for two structural forms of TiO2, namely: anatase and rutile. Applying the calculation relationship (Eq. 3) for particle dimension, it resulted in the average values for anatase (24 nm) and rutile (29 nm) crystallites.

The elemental mapping, Figure 2d, for the silica microfiber decorated with TiO2, evidenced the presence of Ti, Si, and O, as well as the C element. The recorded presence of carbon could be explained by the retention within the material pores and interstices traces of the organic material used during the synthesis stage, although the thorough washing of the obtained material was appropriately performed.

The percentual content in anatase and rutile phases of the titania crystalline nanoparticles was estimated from the XRD pattern (Figure 2c) by the help of the Eqs. (4) and (5). The calculated values indicated the content of 30.8% anatase and 68.9% rutile for the mixed-phase.

The other tetraalkoxy titanate – alcohol systems considered, lead to the formation of titanium oxide as non-adherent nanoparticles (Figure 3a–c).

Thermal analysis of SiO2 microfibers decorated with nanoTiO2 particles. Thermal analysis (TGA and DSC) could provide helpful information regarding the behavior of the SiO2 – TiO2 system and its stability – Supplementary Material Figure S.3. The performed thermal analysis of the quartz fibers decorated with microparticles nanostructured of TiO2 (Figure S.3b) allowed the comparison of the specific characteristics with those of the silica support fibers (Figure S.3a).

The sample of silica microfiber (Figure S.3a) proved to be stable over time, according to the assumption that it does not undergo any transformation. The specimens of silica microfibers decorated with titanium dioxide lost their mass during three stages of heating (Figure S.3b).
UV-Vis –Differential reflectance spectroscopy (DRS) of SiO₂ microfibers decorated with nanoTiO₂ particles. The DRS spectra for the TiO₂ grafted on silica microfiber, nanotitania Degussa P25 (as reference), and initial silica microfiber are presented in Figure 4a.

According to the spectra from Figure 4a the cut-off values obtained were 408 nm for quartz fibers decorated with TiO₂, and 385 nm for the silica microfiber. When the system absorbs radiation with an energy exceeding the band gap of the material, the transfer of electrons from the valence band into the conduction band becomes possible. In Figure 4b it is presented the variation of $(F(R) \cdot h\nu)^2$ versus $E(eV)$ for the direct electronic transition. For studied samples: quartz fibers decorated with TiO₂, silica microfiber, and nanotitania Degussa P25, the band gaps were obtained by extrapolating the linear section to the E axis. The silica microfibers grafted with nanotitania activated with 408 nm radiation showed a band gap of 3.37 eV. This value is lower compared with the band gap values recorded for silica microfiber, 3.61 eV, and TiO₂ Degussa P25, 3.43 eV. The latest band gap values are higher than the value characterizing the obtained material. Therefore, we could consider that the obtained quartz fibers decorated with TiO₂ present a higher degree of crystallinity, and, as a consequence, an improved photocatalytic activity is expected.

Electrokinetic characteristics of SiO₂ microfibers decorated with nanoTiO₂ particles. When used a certain material in suspension, due to the random movement of the particles, it could aggregate if the repulsive forces are not sufficient. The repulsive forces between particles increase as zeta potential increases. In consequence, zeta potentials
allow assessing the stability of a given system. The distribution of potential zeta for the quartz fibers decorated with TiO$_2$ is presented in Figure 4c. The diagram shows that the obtained material has a zeta potential of 22.0 mV. In Supplementary Material, Figure S.4 presents the variation of zeta potential with the pH change for quartz fibers decorated with TiO$_2$. The iso-electric point (IEP) for the studied material calculated from the recorded experimental data was IEP$_{SiO_2-TiO_2}$ = 2.6. Also, the zeta potential is particle-size dependent. Figure 4d presents the particle size distribution. The size distribution by volume is introduced in Supplementary Material Figure S.5. According to literature, the photocatalytic activity of nano-TiO$_2$ depends on the particle size for the same particle surface area. Therefore, it is expected a similar dependence for the silica microfiber – nano-TiO$_2$ composite. The average diameter was ranging between 185.5 nm to 4836 nm underlying the specific deposition of nanotitania onto silica when the nanoparticles agglomerates, forming a nanostructured decoration.

Brunauer-Emmett-Teller (BET) analysis of SiO$_2$ microfibers decorated with nanoTiO$_2$ particles. Figure 5 depicts the N$_2$ adsorption/desorption type IV isotherm and pores size distribution. In Figure 5a, the isotherm describing the material behavior presents a hysteresis loop.
The determined surface area was 101.62 ± 1.38 m²/g, while the pores volume was 0.036 cm³/g. Figure 5b displays the pore size distribution of the obtained material. From the distribution curve, it was determined the average dimension of pores – 2.90 nm.

The surface structure of silica microfibers decorated with nano-TiO₂. The XPS probed the chemical composition of the grafted silica microfibers, Figure 6. The surface analysis of the SiO₂ microfiber grafted with nano-TiO₂ could be explained by the balance between the relatively hydrophobic surface of TiO₂ and the hydrophilic nature of the conditioned SiO₂ surface when OH⁻ amount increased. The presence of the Ti peak on XPS spectra sustains the grafting of TiO₂ nanoparticles on the silica microfibers surface. It was observed that the specific Si peaks became weaker, proving the covering of the microfiber. The carbon C1s presence was due to the synthesis pathway.

The surface of silica microfiber was hydrolyzed, thus being covered with hydroxyl groups, and favoring the titania nanoparticles grafting. The Si atoms are involved in the formation of Si-O-Ti bond, and due to the preferential transfer of electrons from Ti to O, it resulted in larger electron densities for Si. As a consequence, the grafting of the titania nanoparticles leads to a decrease of Si2p binding energy. Much more, at the interface, there is an indication that a bond between Ti and Si was formed, rather than physical adsorption.

Determination of the catalytic characteristics of silica microfibers decorated with TiO₂. In order to determine the photocatalytic properties of the silica microfibers grafted with TiO₂, there was followed up the effect of significant factors on the photodegradation of the cytostatic drug. Key parameters in photocatalysis that were investigated in this study are the catalyst quantity loaded, initial concentration of the analyte, and the influence of irradiation time.

Effect of catalyst amount. For establishing the appropriate catalyst amount, herein quartz fibers decorated with titanium dioxide, it has been followed up the degradation rate for CP versus the catalyst dose for 30 min irradiation (UV) and considering a CP concentration of 7.25 × 10⁻⁵ mol/L. The variation of CP's degradation rate with the increase of the catalyst amount – silica decorated microfibers, is presented in Supplementary Material– Figure S.7.

Effect of irradiation time. The effect of irradiation time was studied following the photocatalytic behavior of [SiO₂-TiO₂] and TiO₂ Degussa P25 over a 30–240 min. range. The dose of the catalyst in both cases was 100 mg/L. Figure 7 introduces the obtained experimental results related to the variation of normalized CP residual concentration with irradiation time.

Effect of initial CP concentration. Table 1 contains the data related to the influence of CP's initial concentration upon the photodegradation efficiency. Irradiation time was set to 30 minutes, while the dose of the catalyst [SiO₂-TiO₂] was 100 mg/L.

Applying the specific reaction rate equation, the obtained data allowed to model the photochemical system behavior through Langmuir – Hinshelwood kinetic.

Discussions

Morphological and structural analysis of SiO₂ microfibers decorated with nanoTiO₂ particles. The SEM analysis performed on the silica microfibers put in evidence a uniform and constant surface aspects (Supplementary – Figure S.1) as well as a smooth surface characterized by 12–13 μm for diameter and 100 μm for length (Supplementary – Figure S.1.c).

For the titanium dioxide microparticles grafting on silica microfibers (Supplementary – Figure S.2.a.), the latest was superficially activated with potassium alkoxides in the corresponding alcohols (methanol, ethanol, propanol, and t-butanol.) The surface's activation creates primary centers on the microfibrils subsequently introduced into the titanium dioxide grafting medium (Supplementary - Figure 2.b).
For grafting the TiO$_2$ onto the silica microfibers' surface, as presented in Section "Procedure for preparing quartz fibers decorated with TiO$_2$", there were four systems considered: tetramethoxy-titanate/methyl alcohol, tetraethoxy-titanate/ethyl alcohol, tetrapropoxy titane propyl alcohol, and tert-butylate titanate/tert-butyl.

Table 1. Initial CP concentration vs. degradation efficiency for silica microfibers – titania decorated catalyst; dose 100 mg/L and irradiation time of 30 min.

| [CP]$_0$ | [CP] | [CP] | η$_{CP}$ |
|----------|-------|-------|----------|
| mg/L     | mol/L | mg/L  | mol/L    | %         |
| 1.04     | 3.98 $\times$ 10$^{-6}$ | 0.46  | 1.76 $\times$ 10$^{-4}$ | 55.77     |
| 2.12     | 8.12 $\times$ 10$^{-6}$ | 0.96  | 3.68 $\times$ 10$^{-4}$ | 54.72     |
| 5.04     | 1.91 $\times$ 10$^{-5}$ | 2.29  | 8.77 $\times$ 10$^{-4}$ | 54.56     |
| 9.83     | 3.77 $\times$ 10$^{-5}$ | 5.23  | 1.88 $\times$ 10$^{-3}$ | 49.95     |
| 18.94    | 7.25 $\times$ 10$^{-5}$ | 11.08 | 4.25 $\times$ 10$^{-3}$ | 41.50     |
| 52.13    | 2.00 $\times$ 10$^{-4}$ | 30.93 | 1.19 $\times$ 10$^{-4}$ | 40.67     |

Figure 6. XPS spectra for the silica microfibers grafted with nanotitania.

Figure 7. Normalized CP residual concentration vs. irradiation time.
alcohol. The tert-butylated-titanate/tert-butyl-alcohol couple was the only system that resulted in adherent grafting and relatively well-distributed titanium dioxide onto the surface of quartz microfibers (Figure 2). As shown in Figure 2a, the quartz microfibres dimension increased towards 14 μm due to the formation of titania nanostructured microparticles of 1–3.5 μm over the smooth silica surface.

The EDX analysis performed on the decorated silica microfibers (Figure 2b) confirmed the formation of titania onto the quartz surface. The XRD analysis showed that from the structural point of view, the titania decoration covering the silica microfibres consists of two forms: anatase and rutile (Figure 2c). The titania nanoparticles are presenting the calculated dimensions of 29 nm (anatase phase) and 33 nm (rutile phase) agglomerate to form microparticles, as evidenced by SEM images (Figure 2a).

Based on the specific properties of each structural phase of TiO₂, there are particular applications like UV filters and photocatalyst for anatase form or materials with magnetic properties and catalyst for rutile form. It is expected that the obtained SiO₂–nanoTiO₂ hybrid material to potentially act as active photocatalyst. The XRD analysis results show the presence of both phases: anatase and rutile wrapping the silica microfibres. Therefore, it was necessary to calculate the ratio between the two structural forms. The higher content of the rutile phase, 68.9%, provided higher stability to the composite system obtained. The presence of both TiO₂ phases could enhance photocatalytic activity. It is known the better photocatalytic behavior of anatase variety. However, the rutile nanoparticles added to anatase nanoparticles could improve the hydroxyl groups’ discharge and the electron transfer and in consequence, the photocatalytic activity.

The other three systems considered for synthesis, tetra-alcoxytitanat – alcohol (see Section 4.3.), lead to the formation of non-adherent titanium dioxide nanoparticles (Figure 3a–c).

**Thermal analysis.** The complete thermal analysis performed on the SiO₂ microfibres – with and without titania decoration, evidenced their stability (Supplementary Material, Figure S.3.a). The residual mass of 99.25% indicates that, eventually, water traces or -OH groups from the surface are lost (Figure S.3.a). The SiO₂–TiO₂ sample showed mass losses over three intervals (Figure S.3.b). The first mass loss of 2.98% occurred from the ambient temperature to 175 °C can be attributed to the removal of water molecules (or another volatile solvent) and weakly -OH linked groups to the particle's surface.

In the 175–340 °C range, the oxidative degradation occurred, and the sample lost 3.77% of its weight. The exothermic effect produced during the thermal process had a maximum decomposition rate at 280.2 °C and an area of 34.61 J/g. This is caused both thermolysis of the amorphous carbon traces remaining from the oxidative degradation. Also, a slight exothermic effect produced during the thermal process had a maximum decomposition rate at 280.2 °C and an area of 34.61 J/g. This is caused both thermolysis of the amorphous carbon traces remaining from the oxidative degradation. Also, a slight exothermic effect produced during the thermal process had a maximum decomposition rate at 280.2 °C and an area of 34.61 J/g. This is caused both thermolysis of the amorphous carbon traces remaining from the oxidative degradation.

In consequence, the hydroxyl group on the surface will dissociate, imparting a negative charge to the surface of the quartz microfibers decorated with TiO₂ due to the formation of TiO⁻ and SiO⁻ species. The nature of the surface hydroxyl from the silica microfibers decorated with the TiO₂ surface is an essential factor in the photocatalytic process. However, the IEP value is low compared with the values reported in the literature. As our material has an IEP lower than 7, it has an acidic character. A higher value for zeta potential could secure a stable suspension, while a low value (<10 mV) leads to particle association and instability. For the silica microfibers decorated with TiO₂, the value of zeta potential was 22 mV. As a high value for ξ assure a maximum dissociation of the functional groups located on surface and particle dispersion, it is expected for the synthesized composite to have a well-dispersed suspension and the surface chemical species will be readily dissociated.

**Electrokinetic studies.** Modifications of zeta potential are generated by adsorption, ionization, or exchanges of chemical species. Thus, when zeta potential (ξ) decreases (its module), then the system is characterized through a high ionic strength, signifying that a significant number of charge carriers are present. In general, when pH decreases, the surface charge becomes more positive, while the pH increase dictates a more negative charge. It is known that when immersing oxides in aqueous solution, the hydroxyl groups of the metallic cations are not altered if the solution has the pH value at the isoelectric point (IEP) of respective oxide. In our case, the IEP for TiO₂ was 4.2 mV, and for silica, microfiber is 2.4 mV. As shown in Supplementary Material Figure S.4. for the obtained quartz fibers decorated with nanoTiO₂, the value of IEP is 2.6 mV. The experimental work was performed at a pH value higher than IEP, in a basic media for synthesis, and towards neutral during the photocatalytic processes.

In consequence, the hydroxyl group on the surface will dissociate, imparting a negative charge to the surface of the quartz microfibers decorated with TiO₂ due to the formation of TiO⁻ and SiO⁻ species. The nature of the surface hydroxyl from the silica microfibers decorated with the TiO₂ surface is an essential factor in the photocatalytic process. However, the IEP value is low compared with the values reported in the literature. As our material has an IEP lower than 7, it has an acidic character. A higher value for zeta potential could secure a stable suspension, while a low value (<10 mV) leads to particle association and instability. For the silica microfibers decorated with TiO₂, the value of zeta potential was 22 mV. As a high value for ξ assure a maximum dissociation of the functional groups located on surface and particle dispersion, it is expected for the synthesized composite to have a well-dispersed suspension and the surface chemical species will be readily dissociated.

**BET analysis of SiO₂ microfibers decorated with nanoTiO₂ particles.** The BET analysis resulted in a type IV isotherm, Figure 5a, that is specific for porous adsorbents with pores in the area of mesopores and macropores, covering the range of 1.5–100 nm. According to the shape of the hysteresis loop, the system exhibits an H4 type isotherm, when the two branches of the loop are almost parallel and horizontal over an extended range of relative pressure, and the geometry of the pore is similar to a narrow slit-shaped structure.

The mesostructured pores appeared from the nanotitania deposition onto the quartz microfibers when voids were formed. In general, an improved photocatalytic activity is conditioned by the high surface area (101.62 ± 1.38 m²/g) and pore volume.

**UV-Vis-DRS analysis.** The results of DRS analysis for silica microfibers grafted with nanotitania highlighted a band gap of 3.37 eV. Such band gap values could be associated with the crystalline characteristics of the material.
that are in good agreement with the X-Ray diffractogram, Figure 2c. When grafting the TiO₂ nanoparticles onto the silica microfibers, for the chemical process:

\[ \text{Ti(OR)}_4 + 4\text{H}_2\text{O} \rightarrow \text{Ti(OH)}_4 + 4\text{R(OR)} \]

an excess of the aqueous solution was used, exceeding the stoichiometric ratio. This situation leads to a nucleophilic reaction between the \( \text{OH}^- \) (from water molecules) and \( \text{OR}^- \) (from alkoxide molecules), thus lowering the possibility to have non-hydralyzed \( \text{OR}^- \) and in consequence, the crystallization of titania is favored.

The titanium oxide variety, rutile, allows a straightforward promotion of electrons from the valence band to the rutile conduction band when exposed to UV excitation, but rapid electron-hole pair recombination accompanies the process. Following the combined processes, there are limited charge carriers that could be involved in various photochemical reactions occurring at the surface level of the oxide. On the other hand, anatase variety that behaves like an indirect band gap semiconductor allows the generation of more holes and electrons available for photo-reactions. In consequence, rutile is not very efficient as a catalyst compared to anatase that presents a substantial amount of electrons and holes reaching its surface. However, there should be specified that such consideration does not take into account any influence of the surface area.

The titania immobilized on the silica microfibers obtained during this work, present both TiO₂ varieties forming a type II heterojunction photocatalyst. Therefore, specific anatase-rutile heterojunctions are present when the rapid mass transfer occurs, and the separation efficacy of hole-electron is satisfactory. Based on the specific values of the conduction and valence band potentials, the obtained composition of the coated titanium oxide allows a two-way charge displacement between anatase and rutile: the generated holes migrate between the valence bands from anatase towards rutile, and the photo-excited electrons move between the conduction bands from rutile into anatase. These charge migrations generate an increase of the electrons and holes that could improve the overall behavior as photocatalyst. As discussed previously, the titanium oxide photocatalyst proved to be efficient in the treatment of wastewater pollutants. The extended applicability of nano-TiO₂ as photocatalyst is limited due to the rapid recombination of electron-hole pair and a large band gap. In the context of the efforts to overcome these drawbacks, the obtained quartz fibers decorated with TiO₂ presents a lower band gap, thus enhancing the photocatalytic activity. In the obtained material, the electrons flux occurs from the conditioned SiO₂ microfiber characterized by a higher band gap (3.61 eV) towards the intrinsically n-type semiconductor TiO₂ that has a lower value of band gap (3.43 eV).

**XPS studies.** The electron density of a chemical species influences the value of its binding energy. A decrease in binding energy implies a higher electron density. As presented in Figure 6, the Ti 2p binding energy for SiO₂ microfiber-covered with nano-TiO₂ decreased to 458.0 eV compared with the literature data. The binding energy indicates a change in the chemical environment and an increase in the electron densities at the Ti level. In literature, it has been reported an increase of the Ti2p binding energy for silica microfibers grafted with TiO₂. The formed Ti-O-Si bond involves the electrons transfer from Ti and Si towards oxygen, and taking into account that Si supplies fewer electrons than Ti, the electron densities level will decrease due to its electrons transfer towards oxygen. The different behavior of our SiO₂-TiO₂ system is due to the distinct synthesis method applied. We could consider that we have predominant Ti-O-Ti bonds. Taking into account that the XPS analysis can provide information about the surface top layers (10 nm depth), and we do not have a thin film deposition of nanotitania, but microparticles nanostructured as resulted from scanning electron microscopy, we means that the detected presence of Si was possible due to incomplete coverage of the silica microfiber with titania nanoparticles. Therefore, to a certain extent, it could be assumed that in any area of SiO₂ microfiber - nano TiO₂ composite surface, only SiO₂ or TiO₂ brings its contribution to the XPS spectrum.

The XPS spectra (Figure 6) displays the shift of the Si 2s peak towards lower binding energy (102.0 eV) compared to the bulk SiO₂ (103.3 eV). This shift of the SiO₂ surface binding energy could be assigned to the lower binding energy of the Si placed at the interface during the TiO₂ grafting onto SiO₂, and to the formation of Ti-O-Si bond replacing Si-O-Si bond. Although the binding energies for Ti 2p peaks put in evidence that the titanium is in its Ti⁴⁺ state, there is a shift towards lower binding energies from the characteristic value for Ti⁴⁺/ TiO₂ bulk of 458.5 eV to 458.0 eV for Ti₂p₁/₂, while the peak for Ti₂p₃/₂ is placed at 463.4 eV. Also, the O1s peak was shifted to lower values, with a maximum at 529.5 eV following an increased negative charge density at O level partially due to a higher degree of covalence of Si-O bond compared to the ionic character of Ti-O bond.

The reduced decrease in the binding energy value for Ti₂p₃/₂ sustains the presence of titanium in its Ti⁴⁺ state. The characteristic Ti2p spin-orbit splitting structure, with a 5.4 eV separation between peaks, was highlighted in Supplementary Material, Figure S6b being in concordance with the data reported in the literature. Such behavior indicates lower binding energy for titanium species compared with the titania bulk. The Ti2p and Si2p bind energies shift towards lower values, indicating higher charge density on Ti and Si following the formation of Ti-O- and Si-O- terminal bonds.

The specific signal for O1s resulted from overlapping the specific bands for both oxides and hydroxyl as well (Supplementary Material, Figure S6a). The values characteristic for the binding energy for O1s in SiO₂ (533.0 eV) and TiO₂ (529.9 eV), as well as the contribution of OH⁻ (531.0 eV), are also well shifted towards lower binding energies, respectively to 529.5 eV for TiO₂-O1s and 530.0 eV for OH⁻ O1s. The exhibited O1s peaks could be ascribed to the surface OH (530.0 eV) and the O²⁻ species of titania (529.5 eV). The success of the material as photocatalyst depends on the reactivity, chemical state, and hydroxyl groups.

Also, we should notice the presence of the Cls signal (at 285.0 eV) corresponding to the C-O in alcohol, signifying that the material morphological structure allowed the accommodation of some traces of organic alcohols used during the material synthesis, although thorough wash of the material was performed.
For the surface ratio of O, Ti, and Si calculated as \( A_O/(A_{Ti} + A_{Si}) \) from XPS data, it was obtained 1.53. It has been reported that for a ratio smaller than 2.00, the surface presents oxygen vacancies\(^5\). The presence of oxygen vacancies generates defects able to act as a trap for holes\(^5\). In consequence, the holes and excited electrons recombination is limited, implying an improved charge transfer.

**Photocatalytic characteristics of silica microfibers decorated with TiO\(_2\).** As shown in section dedicated to the "Effect of catalyst amount," it could be followed-up the variation of cyclophosphamide (CP) degradation rate with the catalyst dose for an initial CP concentration of \( 7.25 \times 10^{-3} \) M, and an irradiation time of 30 min – see Supplementary Material, Figure S.7. The reaction rate increased with the increase of catalyst amount up to \([\text{SiO}_2-\text{TiO}_2]\) = 100 mg/L. The increase of the catalyst surface area available for photocatalysis could explain the observed behavior.

However, a further increase of catalyst amount led to the decrease of reaction rate due to the light scattering effect with a negative impact on photoexcitation efficiency and, therefore, on the CP degradation rate. This trend was also observed in previous studies on CP photocatalytic degradation\(^5\) when there was an optimum catalyst amount established at 100 mg/L. The optimum catalyst dose assures a CP degradation of 41.5% after 30 min of irradiation.

The behavior of silica microfibers decorated with TiO\(_2\) was followed up by comparison with the known catalyst TiO\(_2\) Degussa P25, when equal catalyst doses have been used, namely 100 mg/L. The experimental results, normalized CP residual concentration with irradiation time, presented in Figure 7, were obtained for 30–240 min irradiation time. It has been proved that prolonged irradiation time has a positive effect on CP degradation efficiency. In the case of SiO\(_2\)-TiO\(_2\) catalyst, a CP degradation efficiency of 99.99% was obtained after 30 minutes of irradiation for an initial CP concentration of \( 7.25 \times 10^{-3} \) M. The determined value is satisfactory compared with CP degradation efficiency of 99.04% obtained when the same catalyst dose of TiO\(_2\) Degussa P25 was used in similar experimental conditions. The fact that TiO\(_2\) in the SiO\(_2\)-TiO\(_2\) material is in a much lower amount compared with TiO\(_2\) Degussa P25 suggests that SiO\(_2\)-TiO\(_2\) could be a better photocatalyst.

The composite material obtained takes advantage of the silica substrate presenting corrosion protection, excellent mechanical strength, or high thermal stability. In the meantime, the titanium oxide semiconductor can absorb UV radiation to photo-generate electrons, and the corresponding holes that produce hydroxyl radicals when immersed in aqueous solutions. As a consequence, the photocatalytic activity could be improved due to an increased presence of OH•. Taking into account that the OH• could be produced through the chemical reaction between water and a surface highly acidic, it is expected that the obtained composite would absorb significant amounts of hydroxyl radicals compared to the titanium oxide. This behavior is substantiated by the presence of Si-O-Ti bonds that are acidic, thus conferring a high acidic characteristic to the silica microfibers decorated with nanotitania. Therefore, it is expected that the obtained composite material to have improved photochemical activity.

Kinetic curves for CP degradation when using both SiO\(_2\)-TiO\(_2\) and TiO\(_2\) Degussa P25 were linearized using a pseudo-first-order kinetic, and the constant rates in both cases were calculated for an initial CP concentration of \( 7.25 \times 10^{-3} \) M and catalyst dose of 100 mg/L (Supplementary Material – Figure S.8). The constant rates calculated from the slopes of the graphs were 1.97 \( \times \) 10\(^{-2}\) M/min in the case of TiO\(_2\) Degussa P25 and 1.51 \( \times \) 10\(^{-2}\) M/min in the case of SiO\(_2\)-TiO\(_2\) photocatalyst.

Further studies followed the influence of CP’s initial concentration upon the degradation efficiency (Table 1) when the \([\text{SiO}_2-\text{TiO}_2]\) concentration was 100 mg/L, and the irradiation time applied was 30 min. In the same working conditions, the increase of initial concentration led to the decrease of the degradation efficiency because higher CP levels increase the content of degradation intermediates, which compete with CP for hydroxyl radicals’ consumption.

The Langmuir – Hinshelwood kinetic model can describe CP degradation using SiO\(_2\)-TiO\(_2\) catalyst starting from the following reaction rate Eq. (1):

\[
r_0 = \frac{k_r K_\text{ad}[CP]_0}{1 + K_\text{ad}[CP]_0}
\]

where:
- \( r_0 \): initial CP degradation rate (M min\(^{-1}\))
- \([CP]_0 \): initial CP concentration (M)
- \( k_r \): rate constant for CP photocatalytic degradation (min\(^{-1}\))
- \( K_\text{ad} \): equilibrium constant of CP adsorption–desorption on photocatalyst surface (M\(^{-1}\))

The above-mentioned Eq. (1) could be rearranged as follows Eq. (2):

\[
\frac{1}{r_0} = \frac{k_r K_\text{ad}[CP]_0}{1} + \frac{1}{k_r}
\]

The Eq. (2) allows to plot the relationship: 1/\( r_0 = 1/[CP]_0 \), as linear dependence, in order to calculate the values for the rate constant \( k_r \) and adsorption – desorption equilibrium constant \( K_\text{ad} \). From the linear representation, the intercept gives 1/\( k_r \) while the slope is 1/\( k_r K_\text{ad} \) (Supplementary Material – Figure S.9).

The adsorption–desorption equilibrium constant \( K_\text{ad} \) was 3698 M\(^{-1}\) and the constant rate \( k_r \) was 5.10 \( \times \) 10\(^{-6}\) M min\(^{-1}\). By correlating the experimental data with the Langmuir – Hinshelwood equation results that the degradation of adsorbed CP occurs onto the nano-TiO\(_2\) surface where the photogenerated radicals are also adsorbed.
The values of $k_r$ and $K_{ad}$ obtained for SiO$_2$-TiO$_2$ photocatalyst are similar to those obtained in our previous works for CP degradation using TiO$_2$ in anatase form$^{53}$, which suggests that the new material could be a promising and efficient catalyst for CP’s degradation.

Generally, when referring to the cyclophosphamide’s photooxidation mechanism, it should take into consideration both intermediate and final degradation products (carbon dioxide, ammonium, phosphate, and chloride anion).

In a previous work of our group, based on identified intermediates/degradation by-products$^{54}$, there was proposed a possible alternative for the CP degradation pathway under UV-Vis/silica microfiber decorated nano-TiO$_2$ assisted photocatalysis – Supplementary Material Figure S.10. The photodegradation pathway supposes a direct action on CP. The following processes Eqs. (II) to (XI) for the photo mechanism could apply$^{55}$:

$$\text{SiO}_2@\text{TiO}_2 + h\nu \rightarrow \text{TiO}_2(e_{CB}^-) + \text{TiO}_2(h_{VB}^+)$$  \hspace{1cm} (II)

$$h_{VB}^+ + H_{ads}^+ + \text{HO}_{ads}^- \rightarrow H^+ + \text{HO}$$  \hspace{1cm} (IV)

$$\text{Generation of superoxide anion radicals, } \text{O}_2^-$$

$$\text{(V)}$$

$$\text{Formation of hydroxyl radicals, } \text{HO}$$

$$\text{(VI)}$$

$$\text{Action paths on CP:HO}^* \text{ radicals oxidative action on the CP}$$

$$\text{CP} + \text{HO} \rightarrow \text{H}_2\text{O} + \text{CP} \rightarrow \text{degradation products}$$  \hspace{1cm} (IX)

$$\text{Holes direct oxidative action on CP}$$

$$\text{CP} + h_{VB}^+ \rightarrow \text{CP}^+ (\text{oxidation products})$$  \hspace{1cm} (X)

$$\text{Photodecomposition CP}$$

$$\text{CP}^+ \rightarrow \text{CP photodegradation products}$$  \hspace{1cm} (XI)

By the thin limit between the toxic effect and the targeted efficacy of the cyclophosphamide, and the recorded presence in wastewaters of its toxic metabolites, essential efforts are oriented towards developing new efficient methods to neutralize their toxicity. The new proposed material proved its efficiency for the cyclophosphamide’s photo-oxidation. After 30 min of irradiation, 41.5% cyclophosphamide degraded under UV/silica microfiber decorated TiO$_2$. The Langmuir – Hinshelwood model accurately describes the photo-oxidation reaction. Also, the kinetic data of the photocatalytic process highlighted that the proposed material is a viable solution for advanced photo-oxidation of the chemotherapeutic drug, cyclophosphamide.

Thus, we proved that it is possible to give a superior usage to a valuable waste material – silica microfibers from optical fibers manufacturing industry. The new composite obtained offers an efficient and cheap method for neutralizing the effects of toxic pharmaceutical micropollutants from the wastewaters.
Materials and Methods
All chemicals used, namely: methyl, ethyl, propyl, and t-butyl alcohol, methoxide, ethoxide, propoxide, and potassium t-butoxide, titanium (IV) methoxide, titanium (IV) ethoxide, titanium (IV) propoxide, tetrathyl orthotitanate, and tetrabutyl titanate were from Sigma Aldrich (Merck KGaA, Darmstadt, Germany). The silica microfibers were provided as waste from the processing of the optical glass fibers (Lanxess NV, Belgium). Ultrapure water was used (Millipore, Merck KGaA, Darmstadt, Germany).

For photocatalytic experiments N, N-bis(2-chloroethyl)-1,3,2-oxaza phosphinan-2-amine 2-oxide (cyclophosphamide-CP) (Sigma Aldrich – Merck KGaA, Darmstadt, Germany) and titanium dioxide (Degussa P25/Aeroxide TiO2 P25 – Merck KGaA, Darmstadt, Germany), trifluoroacetic anhydride, sodium chloride and toluene, all from Sigma Aldrich (Merck KGaA, Darmstadt, Germany) have been used.

A Sartorius installation (Sartorius Lab Instruments GmbH&Co, KG., Gottingen, Germany) facilitated the microfiltration of the mixtures obtained by the help of an ultrasonic bath (Elmasonic S10 H, Elma Schmidbauer GmbH, Singen, Germany). The structural investigations, SEM, and elemental analysis, EDX, have been performed on the FE-SEM Hitachi S4500 system (Hitachi High-Technologies Europe GmbH, Krefeld, Germany). The analysis performed on a Netzsch Thermal Analyzer (NETZSCH-Gerätebau GmbH, Selb, Germany) allowed the thermal stability characterization. The structural analysis by X rays diffraction, XRD, has been performed on an equipment Rigaku Ultima IV X-Ray Diffractometer (Rigaku Europe SE, Neu-Isenburg, Germany) using the Cu $K\alpha$ radiation, $\lambda = 1.5406\,\text{Å}$.

Taking into account the specific procedure proposed for nanotitania grafting onto silica microfibers, it is of interest to determine the dimensions of the crystalline nanoparticles. The average size (nm) of nano TiO$_2$ crystallites (D) was estimated through the known Scherrer equation$^{56}$ applied for spherical nanoparticle with a cubic lattice, according to Eq. (3):

$$D = \frac{0.94 \lambda}{\beta \cos \theta}$$

where $\lambda = 0.15406\,\text{nm}$ is the X-ray wavelength of the incident radiation, $\beta$ stands for the full width at half maximum expressed in radians ($= \beta(\text{deg.}) \cdot \frac{\pi}{180}$), $\theta$ is the angle between the diffracted and incident beams (Bragg’s angle of diffraction), considering the Scherrer constant 0.94 that is specific for spherical particles with cubic symmetry$^{57}$.

The ratio of rutile and anatase phase in the grafted nanotitania was evaluated by the help of Spurr-Meyers relationship$^{58}$ as follows Eqs. (4) and (5):

$$\text{% Content in anatase} = \frac{100}{1 + 1.26 \frac{I_R}{I_A}}$$

$\text{Figure 8. Speciation and chemical states on the surface of SiO}_2\text{ microfibers decorated with nano-TiO}_2$. 

[Image: Speciation and chemical states on the surface of SiO$_2$ microfibers decorated with nano-TiO$_2$.]
where the considered maximum intensities $I_a$ for anatase (corresponding to $hkl = (101)$, where $h$, $k$, and $l$ represent the Miller indices) and $I_r$ for rutile phase (for $hkl = (110)$), corresponding to the diffraction angles of 12.63° for anatase maximum peak and 13.71° for rutile, respectively.

Zeta potentials and iso-electric point (IEP) values were determined using Zetasizer UltraPro (Malveren Pannalytical Instruments Ltd., Malveren, UK). It was used electrophoretic light scattering (ELS) for determining the zeta potentials. For hydrodynamic particle size, it was applied the dynamic light scattering (DLS) technique.

The X-ray photoelectron spectroscopy (XPS) analysis performed in order to study the TiO$_2$ grafting onto the silica microfibers surface provided data related to the interactions between the active nano oxide and the supporting silica oxide, and the surface oxygen species. The XPS measurements were run on PHIIBOS 150 ID-DLD system (SPECS Surface Nano Analysis GmbH, Germany) with a monochromatic aluminum anode X-ray source with the energy of 1486.71 eV, $K\alpha$ radiation, as the excitation source. The fixed analyzer transmission mode was applied. For peaks fitting, the SpecsLab Prodigy Version 4.43.2-r73078 (SPECS Surface Nano Analysis GmbH, Germany) was used.

In order to obtain information regarding the optical properties of the obtained SiO$_2$ microfiber grafter with TiO$_2$, there were recorded the reflectance diffusion spectra UV-Vis (UV/DRS), which were obtained using a UV-Vis JASCO V-560 spectrometer (JASCO GmbH, Deutschland) equipped with a diffuse reflectance accessory. From the recorded DRS spectra could be calculated the energy of the band gap, thus assessing the capacity to absorb the UV radiation. The following equation gave the band gap energy Eq. (6):

$$E = \frac{hc}{\lambda}$$

where $h$ is Plank constant – $6.63 \times 10^{-34}$ J·s, $c$ is $3 \times 10^8$ m·s$^{-1}$, $\lambda$ represents the minimum wavelength (cut-off value).

The analysis was run on the material obtained, and for comparison, it was performed the UV/DRS on silica microfiber and nanotitania Degussa P25 as reference.

The electronic characteristics of the synthesized material were analyzed through the Kubelka-Munk remission function as follows Eq. (7):

$$F(R) = \frac{K}{D} = \frac{(1 - R\%)^2}{2 \cdot R\%}$$

where $K$ depends on the incident radiation energy, $D$ represents the dispersion factor, $R\%$ is the diffused reflectance at a particular wavelength normalized to 100.

Then Eq. (8):

$$K = C(E - E_{bg})^n$$

where $C$ stands for a material constant, $E$ is the energy of the incident radiation, $E_{bg}$ represents the band gap, and $n$ takes different values based on the electronic transition type, namely $n = 1/2$ or 2 for permitted direct and indirect transition, respectively.

Thus Eq. (9):

$$F(R) = \frac{c(E - E_{bg})^n}{D}$$

and Eq. (10)

$$E = \frac{hc}{\lambda} = \frac{1236}{\lambda} \text{(eV)}$$

where $h$ is Plank constant – $6.63 \times 10^{-34}$ J·s, $c$ is $3 \times 10^8$ m·s$^{-1}$.

Considering the situation for direct transition, the relationship (9) becomes Eq. (11):

$$F(R)^2 = \left(\frac{c}{D} \right)^2 \left( E - E_{bg} \right)$$

Representing the variation of $(F(R) \cdot h\nu)^2$ versus $E(eV)$ for the direct electronic transition, it could be calculated the band gap value.

The BET surface areas of the synthesized material were determined by the help of N$_2$ adsorption-desorption isotherms measured at 77 K using an ASAP 2020 Plus apparatus (Micromeritics Instruments, USA). N$_2$ adsorption/desorption isotherms allowed the assessment of the specific surface area, pore size distributions, and isotherms for the prepared silica microfiber decorated with nano-titania. The Brunauer–Emmett–Teller (BET) procedure was applied to determine the specific surface area, using the adsorption data in the relative pressure ($P/P_0$) range of 0.05–0.10. The desorption isotherms allowed the calculation of the pore size distribution.

The photocatalytic experiments were designed and run on a Heraeus type UV reactor equipped with a TQ-150-Z3 ($\lambda = 320–550$ nm) mercury lamp (from Heraeus Noblelight GmbH, Hanau, Germany). Agilent
7890 A gas chromatograph (Agilent Technologies Inc., Wilmington, USA) having a column with 5% diphenyl/95% dimethyl-polysiloxane stationary phase allowed to determine the CP's concentration.

**Procedure for preparing quartz fibers decorated with TiO\textsubscript{2}.** Figure S.11, from Supplementary Material introduces the applied method for preparing the silica fibers decorated with TiO\textsubscript{2}.\textsuperscript{29} The experiments were developed using a three-necked flask (250 cm\textsuperscript{3}) fitted with a reflux refigerant. The experimental set-up was accommodated to a mechanical stirrer. 150 cm\textsuperscript{3} of the considered alcohol is added and then stirred with 0.054 moles of silica fibers powder until the transparency of the quartz microfibers occurs. At this particular moment, we introduced 0.054 moles of potassium alcololate. The activation of the fibers' surface is achieved during a half an hour refluxing process of the obtained mixture under continuous stir and heat. The heating stopped after the refluxing is over, but the stirring continued for another four hours. The reaction slurry was filtered on a Sartorius microfilter system, and the excess of potassium alkoxide was removed by repeated washing with the alcohol considered. In parallel, into a 100 cm\textsuperscript{3} conical flask with a plunger stopper, 50 cm\textsuperscript{3} of the considered alcohol and 0.054 mol of titanium tetraalkoxide are ultrasonicated for one hour till the mixture is homogeneous.

As presented in Supplementary Material, Figure S.11, a quantity of 0.010 moles of activated silica microfiber is added, and the complex mixture homogenization is continued for three hours, to suspend the fibers and to complete the TiO\textsubscript{2} gelling reaction onto their surface. A Sartorius microfiltration installation allowed the reaction mixture separation. The obtained silica microfiber decorated with titanium dioxide microparticles nanstructured was subsequently fully characterized.

**Procedure for cyclophosphamide photo-oxidation.** Photocatalytic experiments were performed using a Heraeus type UV reactor equipped with a TQ-150-Z3 (λ = 320–550 nm) mercury lamp. Before irradiation, samples were bubbled with air (50 L/h) in the dark for 30 minutes to avoid holes – electrons recombination.

The following reagents were used in photocatalytic experiments: cyclophosphamide (CP) (Sigma Aldrich), titanium dioxide (Degussa P25), and the synthesized silica microfibers decorated with titanium oxide nanostructured microparticles.

**Analytical methods.** CP concentration was measured using an Agilent 7890A gas chromatograph coupled with an Agilent 240 Ion Trap Mass Detector (GC-MS). As cyclophosphamide has low volatility, derivatization with trifluoroacetic anhydride was employed, forming N-tri fluoroacetyl- cyclophosphamide. Samples from 10 to 400 mL of aqueous samples were treated with sodium chloride (0.1 g NaCl/mL of the sample) at pH 10 buffer (0.5 mL buffer/mL of the sample) and then extracted with toluene. 1 mL extract was transferred in 2 mL GC vial and treated with 20 μL of trifluoroacetic anhydride for derivatization. After sealing and vigorously shaking, vials were heated at 70 °C for 2 h. Subsequently to derivatization, 2 μL of toluene solution was analyzed by GC-MS. A column with 5% diphenyl/95% dimethyl-polysiloxane stationary phase (30 m, 0.25 mm ID, 0.25 μm film) was used for separation of trifluoroacetylated derivatives (RT: 9.88 min for IF derivative and RT:10.38 min for CF derivative). The experiments have been performed under specific working conditions for GC and mass detector, as follows: (a) GC conditions: splitless injection with injector temperature 250 °C, flow rate: 1 mL/min; carrier gas: helium 6.0; temperature ramp employed: 70 °C (1 min), 20 °C/min to 280, hold 2.5 min; (b) mass detector conditions: Ion Trap: 120 °C; Manifold: 50 °C; Transfer line: 250 °C; Ion source: 220 °C; Scan mode: single ion monitoring (quantified ion: m/z= 307 and qualifier ion: m/z= 309).

**Ethical statement.** This work does not contain studies with human participants or animals performed by any of the authors.

Received: 23 September 2019; Accepted: 17 December 2019; Published online: 10 January 2020

**References**

1. Kim, E. J. Thorn-like TiO\textsubscript{2} nanoarrays with broad spectrum antimicrobial activity through physical puncture and photocatalytic action. *Sci. Rep.* 9 (2019).

2. Dalod, A. R. M., Henriksen, L., Grande, T. & Einarsrud, M. A. Functionalized TiO\textsubscript{2} nanoparticles by single-step hydrothermal synthesis: The role of the silane coupling agents. *Beilstein J. Nanotechnol.* https://doi.org/10.3762/bjnano.8.33 (2017).

3. Anastasescu, C. et al. Tubular and spherical SiO\textsubscript{2} obtained by sol gel method for lipase immobilization and enzymatic activity. *Molecules*, https://doi.org/10.3390/molecules23061362. (2018).

4. Almasian, A., Chizari Fard, G. & Maleknia, L. Fabrication of hollow and nonhollow SiO\textsubscript{2} nanofibers for removal of cationic dyes from aqueous solutions. *Environ. Prog. Sustain. Energy*, https://doi.org/10.1002/ep.12584 (2017).

5. Polyakov, B. et al. Manipulation of nanoparticles of different shapes inside a scanning electron microscope. *Beilstein J. Nanotechnol.*, https://doi.org/10.3762/bjnano.5.13 (2014).

6. Schumann, E., Hübner, R., Grenzer, J., Gemming, S. & Krause, M. Percolated Si:SiO\textsubscript{2} Nanocomposites: Oven- vs. Millisecond Laser-Induced Crystalization of SiOx Thin Films. *Nanomaterials*, https://doi.org/10.3390/nano8070525. (2018).

7. Nechifor, A. C., Tutu, E. E., Ivan, A., Danculescu, V. & Sava, S. Microstructured adsorbent material: Silica dioxide – polyaniline for retaining aniline and chromium ions. *Optoelectron. Adv. Mater. Rapid Commun.*, 7, 145–149 (2013).

8. Sung, S. D. et al. 50 nm sized spherical TiO\textsubscript{2} nanocrystals for highly efficient mesoscopic perovskite solar cells. *Nanoscale*, https://doi.org/10.1039/c5nr01364a (2015).

9. Hosseini, A., Kumar, P., Mahdi, N., Zhang, Y. & Shankar, K. All-solid-state formation of titania nanotube arrays and their application in photoelectrochemical water splitting. *J. Mater. Sci. Mater. Electron.*, https://doi.org/10.1007/s10854-018-9752-2. (2018).

10. Valavarthi, K. et al. Radiative and Non-Radiative Recombination Pathways in Mixed-Phase TiO\textsubscript{2} Nanotubes for PEC Water-Splitting. *Catalysts*, https://doi.org/10.3390/catal9020204. (2019).

11. Caratão, B., Carneiro, E., Sá, P., Almeida, B. & Carvalho, S. Properties of Electrospun TiO\textsubscript{2} Nanofibers. *Sci. Rep.*, https://doi.org/10.1038/s41598-019-56836-7. (2019).

12. Seu-Run Kim, S. R., Imran Ali, I. & Kim, J. O. Phenol degradation using an anodized graphene-doped TiO\textsubscript{2} nanotube composite under visible light. *Appl. Surf. Sci.*, 477, 71–78 (2019).
13. Schneider, J. et al. Understanding TiO$_2$ photocatalysis: Mechanisms and materials. *Chem. Rev.*, https://doi.org/10.1021/cr5001892 (2014).

14. Ohno, T. et al. Preparation of 5-doped TiO$_2$ photocatalysts and their photocatalytic activities under visible light. *Appl. Catal. A Gen.*, https://doi.org/10.1016/j.apcata.2004.01.007 (2004).

15. Shen, S. et al. Titanium dioxide nanostructures for photoelectrochemical applications. *Progress in Materials Science*, https://doi.org/10.1016/j.pmatsci.2018.07.006 (2018).

16. Fischer, K. et al. Synthesis of High Crystalline TiO$_2$ Nanoparticles on a Polymer Membrane to Degrade Pollutants from Water. *Catalysts.*, https://doi.org/10.3390/catal9060936 (2018).

17. Tavakolmoghadam, M., Mohammadi, T., Hemmati, M. & Naeimpour, F. Surface modification of PVD films membranes by sputtered TiO$_2$: fouling reduction potential in membrane bioreactors. *Desaln. Water Treat.*, https://doi.org/10.1016/j.memsci.2014.984635 (2016).

18. Totto, E. E., Nechipor, A. C., Nechipor, G., Aboul-Enein, H. Y. & Cristache, C. M. Poly(methyl methacrylate) with TiO$_2$ nanoparticles inclusion for stereolithographic complete denture manufacturing - the future in dental care for elderly edentulous patients? *J. Dent.*, 59, 68–77 (2017).

19. Leandro, M. J., Edwards, J. C. W. & Cambridge, G. Clinical outcome in 22 patients with rheumatoid arthritis treated with B lymphocyte depletion. *Ann. Rheum. Dis.*, 833–88 (2002).

20. Shah, K. M., Ohri, A. J. & A., U. A Randomized Controlled Trial of Intravenous versus Oral Cyclophosphamide in Steroid-resistant Nephrotic Syndrome in Children. *Indian J. Nephrol.*, 27, 430–437 (2017).

21. Mok, C. C. Con: Cyclophosphamide for the treatment of lupus nephritis. *Nephrol Dial Transplant.*, 31, 1053–1057 (2016).

22. Ryerson, C. J. & Denton, C. P. The fine line between success and failure in scleroderma lung fibrosis trials. *Am. J. Respir. Crit. Care Med.*, 197, 553 (2018).

23. Buerge, J. I., Buser, H. R., Poiger, T. & Müller, M. D. Occurrence and fate of the cystostatic drugs cyclophosphamide and ifosfamide in wastewater and surface waters. *Environ. Sci. Technol.*, https://doi.org/10.1021/acs.est.0e00940 (2006).

24. Webb, S., Ternes, T., Gibert, M. & Olejniczak, K. Indirect human exposure to pharmaceuticals via drinking water. *Toxicol. Lett.*, https://doi.org/10.1016/S0378-4274(07)00017-3 (2003).

25. Sarder, A., Rabbani, M. G., Chowdhury, A. S. M. H. K. & Sobhani, M.-E. Molecular Basis of Drug Interactions of Methotrexate, Cyclomethoxantrone and 5-Fluorouracil as Chemotherapeutic Agents in Cancer. *Biomed. Res. Ther.*, https://doi.org/10.7603/s40730-0150005-1 (2016).

26. Tsai-Turton, M., Luong, B. T., Tan, Y. & Luderer, U. Cyclophosphamide-induced apoptosis in COV434 human granulosa cells involves oxidative stress and glutathione depletion. *Toxicol. Sci.*, https://doi.org/10.1093/toxsci/kfm087 (2007).

27. Sies, H. Hydrogen peroxide as a central redox signaling molecule in physiological oxidative stress: Oxidative eustress. *Redox Biology*, https://doi.org/10.1016/j.redox.2016.12.035 (2017).

28. Kot-Wasiak, A., Jakimiska, A. & liwka-Kaszyńska, M. Occurrence and seasonal variations of 25 pharmaceutical residues in wastewater and drinking water treatment plants. *Environ. Monit. Assess.*, 188, 661 (2016).

29. Directive 2008/98/EC of the European Parliament and of the Council of 19 November 2008 on waste and Directive (EU) 2018/851 of the European Parliament and of the Council of 30 May 2018 amending Directive 2008/98/EC on waste, http://data.europa.eu/eli/dir/2018 (2018).

30. Lai, W. W.-P., Chuang, Y.-C. & Lin, A. Y.-C. The effects and the toxicity increases caused by bicarbonate, chloride, and other water components during the UV/TiO$_2$ degradation of ozonaphosphorin drugs. *Environ. Sci. Pollut. Res.*, 24, 14595–14604 (2017).

31. Franquet-Griell, H., Medina, A., Sans, C. & Lacorte, S. Biological and photochemical degradation of cytostatic drugs under laboratory conditions. *J. Hazard. Mater.*, 319–328 (2017).

32. Schulte-Oehlmann, U., Oehlmann, J. & Püttmann, W. Humanpharmakawirkstoffe in der Umwelt: Einträge, Vorkommen und der Versuch einer BestandsaufnahmePharmaceutical agents in the environment, their entries, occurrence, and an inventory of them. *Umweltwissenschaften und Schadstoff-forsch.*, https://doi.org/10.1065/wws2007.07.202 (2009).

33. Kümmner, K. & Al-Ahmad, A. Estimation of the cancer risk to humans resulting from the presence of cyclophosphamide and ifosfamide in surface water. *Environ. Sci. Pollut. Res.*, https://doi.org/10.1007/s11356-009-0195-4 (2010).

34. Zhang, X., Xiao, Y., Zhang, J., Chang, W. C. & Lim, T.-T. Degradation of cyclophosphamide and 5-fluorouracil in water using UV and UV/H$_2$O$_2$: Kinetics investigation, pathways and energetic analysis. *J. Environ. Chem. Eng.*, 5, 1133–1139 (2017).

35. Hromadko, L., Koudelková, E., Bulánek, R. & Macák, J. M. SiO$_2$ Fibers by Centrifugal Spinning with Excellent Textural Properties. *Adv. Mater.*, 319–328 (2017).

36. Hanaor, D. A. H. & Sorrell, C. C. Sand supported mixed-phase TiO$_2$ photocatalysts for water decontamination applications. *Adma.*, https://doi.org/10.1007/s11356-009-0195-4 (2010).

37. Adma.201601694 (2017).

38. Role of Surface Area, Primary Particle Size, and Crystal Phase on Titanium Dioxide Nanoparticle Dispersion Properties. *Nanoscale Res Lett.*, 6, 27 (2011).

39. Li, Q. et al. Antimicrobial nanomaterials for water disinfection and microbial control: potential applications and implications. *Water Research.*, 42, 4591–4602 (2008).

40. Lowe, J., Yu, J., Jaroniec, M., Wageh, S. & A-Ghamdi, A. A. Heterojunction Photocatalysts. *Adv. Mater.*, https://doi.org/10.1002/adma.201601694 (2017).

41. Totto, E. E. et al. On physical and chemical characteristics of Poly(methylmethacrylate) nanocomposites for dental applications. *I. Mater. Plast.*, 54, 666–672 (2017).

42. Totto, E., Segal, E. & Covington, A. K. On the thermal behaviour of some polyimide membranes. *J. Therm. Anal. Calorim.*, 52, 383–391 (1998).

43. Hanor, D. A. H. & Sorrell, C. C. Sand supported mixed-phase TiO$_2$ photocatalysts for water decontamination applications. *Adv. Eng. Mater.*, https://doi.org/10.1002/adem.201300259 (2014).

44. Xing Wang, X. et al. Highly enhanced photocatalytic performance of TiO$_2$ nanosheets through constructing TiO$_2$/TiO$_2$ quantum dots homojunction. *App. Surf. Sci.*, 499, 9–15 (2018).

45. Zhu, X. et al. The effect of heat treatment on the anatase-rutile phase transformation and photocatalytic activity of Sn-doped TiO$_2$ nanomaterials. *Rsc Adv.*, https://doi.org/10.1039/cra00766q (2018).

46. Krishnan, P. et al. Characterization of photocatalytic TiO$_2$ powder under varied environments using near ambient pressure X-ray photoelectron spectroscopy. *Sci.Rep.*, 7, 43298 (2017).

47. Moulder, J. F., Stickle, W., Sobol, P. E. & Bomben, K. D. edited by J. C. *Handbook of X-Ray Photoelectron Spectroscopy*. Perkin-Elmer, US (1992).

48. Tan, R. et al. Theoretical study of the adsorption characteristics and the environmental influence of ornidazole on the surface of photocatalyst TiO$_2$. *Sci.Rep.*, 9 (2019).

49. Meng, A., Zhang, L., Cheng, R. & Yu, J. Photocatalysis. *Adv. Mater.*, 31, 1807660 (2019).

50. Liu, Z., Zhang, X., Murakami, T. & Fujishima, A. A Sol-gel SiO$_2$/TiO$_2$ bilayer films with self-cleaning and antireflection properties. *Sol. Energy Mater. Sol. Cells.*, 92, 1434–1438 (2008).

51. Sachs, M. et al. Effect of oxygen deficiency on the excited state kinetics of WO3 and implications for photocatalysis. *Chem. Sci.*, 10, 5667–5677 (2019).
Acknowledgements
The authors acknowledge the support by a grant of the Romanian National Authority for Scientific Research and Innovation, CCCDI-UEFISCDI, project number 39/2018 COFUND-MANUNET III-HAMELDENT, within PNCDI III. Also, E.E.T. and M.E.C. acknowledge the support of COST Action NutRedOx-CA16112 supported by COST (European Cooperation in Science and Technology) that made possible their co-operation.

Author contributions
Concept, A.C.N., E.E.T.; Funding acquisition, E.E.T., G.N., A.C.N.; Methodology, A.C.N., G.N., I.I., L.C., M.E.C.; Investigation, E.E.T., G.N., I.I., A.C.N., L.C., A.M.C.; Resources, E.E.T., G.N.; Data analysis and interpretation, E.E.T., A.C.N., M.E.C.; Writing—original draft preparation, G.N., E.E.T.; Writing, review and editing, E.E.T.

Competing interests
The authors declare no competing interests.

Additional information
Supplementary information is available for this paper at https://doi.org/10.1038/s41598-019-56836-7.

Correspondence and requests for materials should be addressed to E.E.T.

Reprints and permissions information is available at www.nature.com/reprints.

Publisher's note Springer Nature remains neutral with regard to jurisdictional claims in published maps and institutional affiliations.

Open Access This article is licensed under a Creative Commons Attribution 4.0 International License, which permits use, sharing, adaptation, distribution and reproduction in any medium or format, as long as you give appropriate credit to the original author(s) and the source, provide a link to the Creative Commons license, and indicate if changes were made. The images or other third party material in this article are included in the article's Creative Commons license, unless indicated otherwise in a credit line to the material. If material is not included in the article’s Creative Commons license and your intended use is not permitted by statutory regulation or exceeds the permitted use, you will need to obtain permission directly from the copyright holder. To view a copy of this license, visit http://creativecommons.org/licenses/by/4.0/.

© The Author(s) 2020