Assessment of Changes in Land Use/Land Cover and Land Surface Temperatures and Their Impact on Surface Urban Heat Island Phenomena in the Kathmandu Valley (1988–2018)

Md. Omar Sarif 1, Bhagawat Rimal 2,3,*, and Nigel E. Stork 4

1 Geographic Information System (GIS) Cell, Motilal Nehru National Institute of Technology Allahabad, Prayagraj 211004, India; rgi1606@mnnit.ac.in or mdomarsarif@gmail.com
2 College of Applied Sciences (CAS)-Nepal, Tribhuvan University, Kathmandu 44613, Nepal; bhagawatrimal@gmail.com
3 The State Key Laboratory of Remote Sensing Science, Institute of Remote Sensing and Digital Earth, Chinese Academy of Sciences, Beijing 100101, China
4 Environmental Future Research Institute, Griffith School of Environment, Nathan Campus, Griffith University, 170, Kessels Road, Nathan, QLD 4111, Australia; nigel.stork@griffith.edu.au

* Correspondence: bhagawat@radi.ac.cn

Received: 28 October 2020; Accepted: 23 November 2020; Published: 6 December 2020

Abstract: More than half of the world’s populations now live in rapidly expanding urban and its surrounding areas. The consequences for Land Use/Land Cover (LULC) dynamics and Surface Urban Heat Island (SUHI) phenomena are poorly understood for many new cities. We explore this issue and their inter-relationship in the Kathmandu Valley, an area of roughly 694 km², at decadal intervals using April (summer) Landsat images of 1988, 1998, 2008, and 2018. LULC assessment was made using the Support Vector Machine algorithm. In the Kathmandu Valley, most land is either natural vegetation or agricultural land but in the study period there was a rapid expansion of impervious surfaces in urban areas. Impervious surfaces (IL) grew by 113.44 km² (16.34% of total area), natural vegetation (VL) by 6.07 km² (0.87% of total area), resulting in the loss of 118.29 km² area from agricultural land (17.03% of total area) during 1988–2018. At the same time, the average land surface temperature (LST) increased by nearly 5–7 °C in the city and nearly 3–5 °C at the city boundary. For different LULC classes, the highest mean LST increase during 1988–2018 was 7.11 °C for IL with the lowest being 3.18 °C for VL although there were some fluctuations during this time period. While open land only occupies a small proportion of the landscape, it usually had higher mean LST than all other LULC classes. There was a negative relationship both between LST and Normal Difference Vegetation Index (NDVI) and LST and Normal Difference Moisture Index (NDMI), respectively, and a positive relationship between LST and Normal Difference Built-up Index (NDBI). The result of an urban–rural gradient analysis showed there was sharp decrease of mean LST from the city center outwards to about 15 kms because the NDVI also sharply increased, especially in 2008 and 2018, which clearly shows a surface urban heat island effect. Further from the city center, around 20–25 kms, mean LST increased due to increased agriculture activity. The population of Kathmandu Valley was 2.88 million in 2016 and if the growth trend continues then it is predicted to reach 3.85 million by 2035. Consequently, to avoid the critical effects of increasing SUHI in Kathmandu it is essential to improve urban planning including the implementation of green city technologies.

Keywords: Land Use/Land Cover; NDBI; NDVI; NDMI; SUHI; urban–rural gradient; Kathmandu Valley; Nepal
1. Introduction

People increasingly prefer to live in urban and surrounding areas since these are perceived to offer improved opportunities for employment, good healthcare, and higher level of education [1–3]. As a result, the global urban population has increased to 55% by 2018, and is predicted to increase further to 68% by 2050 [4]. Such growth in the urban population has resulted in increased spread of built up urban areas which have directly or indirectly altered the dynamics of land use/land cover (LULC) at large scales [1,5,6]. Globally, land surface air temperature has increased by 1.53 °C and the global mean surface (ocean and land) air temperature by 0.87 °C since the pre-industrial period (1850–1900) [7]. The dramatic rise in urbanization has negative environmental consequences especially increases in Surface Urban Heat Island (SUHI) effects which severely impact human health, air quality, and climate change, all of which are major topics of global concern [8,9]. The observed higher land surface temperatures (LST) over city centers compared to surrounding areas, the so-called SUHI, results largely from the conversion of natural spaces into built-up areas with largely impervious surfaces [8].

Understanding LULC has become increasingly important to help understand aspects of urban dynamics such as geography, morphology, ecology, and sustainability, with the goals of strengthening concepts of land use patterns, urban intensity, urban diversity, UHI phenomena, among other factors [9]. LULC of urban improvement and urban expansion have been explored widely in both developed and developing nations with spatial expansion appearing to be more complicated in developing nations [10]. Assessment of LULC in some cities has been particularly influential in aiding future urban planning, transport network development, economic prosperity growth, policy development, and environmental enhancement [10–12].

The surfaces of city landscapes have properties that lead to a wide variety of behaviors with respect to phenomena such as electromagnetic radiation absorbance or evaporation, longwave radiation, and prevailing winds, resulting in increased heat discharge [13]. The physical surface of a city comprises different materials such as asphalts, gravels, and stones as well as other building material surfaces, which enhance the sensitivity and lower evapotranspiration in the city with consequent effects on city climate [14,15]. This leads to higher absorption of sun radiation, higher retention of infrared radiation in street canyons, higher uptake, and delayed release of heat from buildings, higher proportion of absorbed sun radiation than latent heat forms, and higher release of latent heat from combustion fuels for industrial processing, urban transport, and domestic space heating [16]. As a result, core urban spaces experience warmer weather than their surrounding areas, which leads to the creation of SUHI. This SUHI phenomenon drastically transforms the local climate with adverse effects on plants as well as animals including humans [10,17]. SUHI has rapidly increased due to the large-scale increase in areas of impervious surfaces in the process of urbanization [18,19]. As a result, SUHI is of increasing concern for urban planners, health authorities, urban investors, policy makers, ecologists, and others due to its adverse effects on various aspects of the environment such as rainfall, temperature, air quality, energy balance, and carbon storage [20,21].

To help mitigate the effects of SUHI, the spatial dynamics of city landscapes need to be assessed [13]. Plans for improving cities should include improved design of building structures, including the use of environmentally friendly materials to minimize thermal energy storage, excessive impermeability, and solar radiance absorbance. Developing sound mitigation strategies will help ‘green’ cities, such as improving the energy balance through the use of cool and green roofs, increased urban tree plantings, and the development of gardens [22]. To aid the development of these strategies there is a need for spatial quantification of LULC and SUHI dynamics using Remote Sensing since these tools provide cost-effective and reliable data [23–27]. Elsewhere, assessment of LULC dynamics has provided land-resource information which has aided sustainable development [28–31]. Recently, the SUHI phenomena [32,33], the interrelationship between LST vs. Normal Difference Vegetation Index (NDVI) [34–36], LST vs. Normal Difference Built-up Index (NDBI) [37,38] and LST vs. Normal Difference Moisture Index (NDMI) [39] have been assessed in several city landscapes using Pearson’s
correlation coefficients [40,41]. These studies confirm that impervious areas increase SUHI and increasing vegetation reduces this effect.

Studies of the growing urban history of Kathmandu in Nepal date back to the late 1950s [42,43] and showed that urban areas have expanded rapidly since this time. While the processes of urbanization and LULC change in Kathmandu has been previously explored [43–47], the role of LULC in LST intensification is not understood. Here we address this knowledge gap assessing the spatiotemporal dynamics of LULC and their influence on LST intensification, creation of SUHI, and the relationship of NDVI and NDBI with LST using Pearson’s correlation coefficient. We specifically examine these issues in the geopolitical area of Kathmandu Valley, which includes the districts of Bhaktapur, Kathmandu, and Lalitpur [42]. We used Landsat datasets (1988–2018) at decadal intervals selecting dry, summer season (April) data [48]. This study area was selected because of its location in the Himalayas, its heterogeneous physiographic diversity, and because it is one of the most rapidly expanding cities in South Asia [43,45]. Considerable LULC change has taken place as urban areas have expanded from the city center altering the dynamics of land-resource prospects and this analysis is essential for future sustainable development in the Kathmandu Valley.

We explore the following: (a) the spatiotemporal dynamics of LULC and their influence on intensification of SUHI for the Kathmandu Valley; (b) for each distinct LULC class we assess the mean LST along with their influence on the city landscape; (c) the relationship between LST and indicators of LULC (NDBI, NDVI, and NDMI) to aid assessment of the relative importance of impervious space and vegetation space with respect to LST in the Kathmandu Valley; and d) LST, NDVI, and NDBI on an urban–rural gradient from the city center to the urban periphery.

2. Materials and Methods

2.1. Study Area

Kathmandu Valley is situated in central Nepal and comprises 694.27 km$^2$, which includes three government districts: Bhaktapur, Lalitpur, and Kathmandu [47], and is located between 27°31’ to 27°50’ North, and between 85°11’ to 85°34’ East in the Himalaya mountains (Figure 1). It had a population of 1.11 million in 1991 which more than doubled to 2.52 million by 2016 [49]. Elevation ranges from 837 m to 2723 m, with the central part of the valley ranging from 1200 m to 1500 m. It has a dry-winter humid subtropical climate (Cwa) [50]. The valley has been created by the Bagmati River [43] which flows from north-east to south-west. Mean annual temperature and precipitation are 18.1 °C and 1407 mm, respectively [47].
2.2. Data

For our study, LULC data were based on images for individual days in April (early summer) at decadal intervals using Landsat TM and Landsat 8 OLI/TIRS at four different time-points (1988, 1998, 2008, 2018) which were retrieved from the USGS website (https://earthexplorer.usgs.gov) (Table 1). These data were also used to create LST based on their thermal bands (Band-6 of Landsat 5-TM, and Band-10 of Landsat 8-TIRS) and NDVI, NDBI, and NDMI based on optical and near-infrared bands (Band-3, 4, and 5 of Landsat 5-TM, and Band-4, 5, and 6 of Landsat 8-OLI), respectively. For the validation of weather conditions, we have explored NASA’s POWER project data on selected sample location for our study area and we have found that weather conditions of our selected time points are satisfactory (please refer to supplementary excel file).

Topographical data at the scale of 1:25,000 [51] and Google Earth images of the study area also were used. To overcome the atmospheric error in satellite images, they were pre-processed, including atmospheric and radiometric correction. Decadal intervals were chosen to determine growth in urban expansion and its influence in the transformation of LST, as it has been suggested as an optimal time period to measure LULC change, as well as SUHI development [10]. The Global Digital Elevation Model (GDEM) of ASTER was used to depict the topography of Kathmandu valley and to help understand its physiographic dynamics. We recognize the fact that just selecting single days to represent a decadal point in time does not address the problem of variation in LST between days or even between years but we consider our data to be a starting point for further analysis that may later address such issues.
Table 1. Description of the Landsat datasets used for the study area, Kathmandu Valley.

| Sensor      | Path/Row | Resolution | Acquisition Date | Time (GMT) | Constants of Thermal Conversion | Source                                      |
|-------------|----------|------------|------------------|------------|-------------------------------|---------------------------------------------|
| Landsat-5   | 141/41   | 30 m       | 3 April 1988     | 04:18:24   | 607.76 (Band 6) 1260.56 (Band 6) | United States Geological Survey (USGS) web portal (https://earthexplorer.usgs.gov/) |
| Landsat-8   |          |            | 15 April 1998    | 04:25:19   | 607.76 (Band 6) 1260.56 (Band 6) |                                             |
|             |          |            | 26 April 2008    | 04:37:22   | 607.76 (Band 6) 1260.56 (Band 6) |                                             |
| Aster       |          |            | 22 April 2018    | 04:47:35   | 774.8853 (Band 10) 1321.0789 (Band 10) |                                             |

2.3. Retrieval of LULC

We used freely available terrain corrected Landsat (Level 1T) datasets with best cloud-free data (less than 10% cloud cover) of UTM zone 45N. Our area of interest (AOI) had zero numbers of spiked digital numbers (DN) due to the availability of cloud free imageries and hence, we did not need to do masking and exclusion for correction of co-registration, cloud, cloud shadow, and gap-filling. Radiometric and geometric corrections are part of the pre-processing function of satellite images. In this study, we used ENVI software to carry out the atmospheric image correction process. Following this the DN values of images were converted into radiance values. All images were further verified for their accuracy and the root mean square (RMS) of the geometric rectification of less than 15 m (0.5) pixels was accepted. The Flash Line-of-sight Atmospheric Analysis of Spectral Hypercubes (FLAASH) model was used for atmospheric correction and LULC data were extracted using ENVI software [12,46].

There are multiple image classification algorithms, such as support vector machine (SVM), maximum likelihood classifier (MLC), and decision tree (DT) [52]. In this study we chose to use SVM for the classification of images as it is flexible, uses a non-parametric approach and is widely used for the extraction of LULC data bases [53]. Further, according to some authors, SVM has a higher level of accuracy than MLC [54]. Similarly, others have shown SVM performed better in the assessment of land cover changes and urban development than MLC and DT [55]. SVM can be grouped into four kernels function, such as linear, polynomial, radial function, and sigmoid [56]. The Radial Basic Function kernel was used at the time of extraction of different LULC classes as it usually provides better results than other machine learning [57]. Detailed field visits were conducted to ground-proof our analyses and five main LULC classes were identified based on the Anderson Classification Scheme at Level-I [58]: Impervious land (IL), Agriculture land (AL), Vegetation Land (VL), Open Land (BL), and Water Body (WB) (Table 1).

Accuracy assessment is essential for land cover data developed from remote sensing technology [59,60]. Overall accuracy (OA), User’s accuracy (UA), and producer’s accuracy (PA) assessment were assessed based on the field reference information. A topographical map of the study area for the scale of 1:25000 was used as developed by the Survey Department of Nepal in 1995 [51]. Similarly, Google Earth images for multiple dates were used in the assessment. OA represents what proportion of references sites were mapped correctly whereas UA is defined as the accuracy from the knowledge of a map user, not the map maker. PA is defined as the accuracy of a map from the knowledge of the map maker (the producer) [60].

Accuracy assessment reports were generated for each class of LULC using 200 stratified random samples points (total 1000 random points) and an error matrix for each time point was created to allow accuracy assessments [61]. UA, PA, OA, and Kappa coefficients were calculated based on the error matrix for each time point (1988–2018).UA, PA, OA, and Kappa coefficients were calculated based on
error matrix of respective time points (1988–2018). To optimize classification accuracy: User, Producer, Overall accuracy, and Kappa coefficients were estimated using Equations (1)–(4) [10,61].

\[
\text{User Accuracy} = \left\{ \frac{\sum_{\omega} \theta(\omega) \times 100}{\sum_{\omega} \theta(\omega)} \right\}
\]

(1)

\[
\text{Producer Accuracy} = \left\{ \frac{\sum_{\phi} \theta(\phi) \times 100}{\sum_{\phi} \theta(\phi)} \right\}
\]

(2)

\[
\text{Overall Accuracy} = \left\{ \frac{\sum_{\beta} \sum_{\gamma} \times 100}{\sum_{\beta} \sum_{\gamma}} \right\}
\]

(3)

\[
\text{Kappa Coefficients} = \frac{N \sum_{i=1}^{r} X_{ii} - \sum_{i=1}^{r} (X_{i+} \times X_{+i})}{N^2 - \sum_{i=1}^{r} (X_{i+} \times X_{+i})}
\]

(4)

where, \( \epsilon \) defines corrected classified pixels (CCP) (category); \( \theta(\omega) \) defines classified pixels (CP) in that category (row total (RT)); \( \theta(\phi) \) defines CP in that category (column total (CT)); \( \beta \) defines CCP (diagonal); \( \gamma \) defines classified reference pixels in that category; \( N \) defines total samples; \( r \) defines number of rows error matrix (EM); \( X_{ii} \) defines total corrected samples in \( i^{th} \) row and column; \( X_{i+} \) defines RT; \( X_{+i} \) defines CT.

2.4. Land Surface Temperature (LST) Retrieval

2.4.1. Land Surface Emissivity

Land surface emissivity (LSE) is an important parameter to calculate LST [62,63]. NDVI threshold (NDVI\text{THR}) method was used to calculate LSE because it can differentiate pixels of vegetation, water, and soil significantly [64]. It was calculated using Equation (5).

\[
\epsilon_{sv} = \epsilon_v P_V + \epsilon_s (1 - P_V) + C
\]

(5)

where, \( \epsilon_{sv} \) is emissivity of soil and vegetation; \( \epsilon_v \) is emissivity of vegetation; \( \epsilon_s \) is emissivity of soil; \( P_V \) is proportion of vegetation (using Equation (6)); \( C \) is defining constant for surface characteristics (using Equation (7)).

\[
P_V = \left[ \frac{NDVI - NDVI_b}{NDVI_v - NDVI_b} \right]^2
\]

(6)

where, \( NDVI \) is estimated using Equation (19) in Section 2.5; \( NDVI_b \) is NDVI of pure soil; \( NDVI_v \) is NDVI of pure vegetation.

\[
C = (1 - \epsilon_s) \epsilon_v F (1 - P_V)
\]

(7)

where, \( F \) is geometric factor (it depends on surface geometry, commonly considered as 0.55 [64,65]).

Skokovic et al. [66], Sorbino et al. [64], Sekertekin and Bonafoni [65] calculated LSE using NDVI\text{THR} values considering three distinct cases shown in Equation (8). The first case had NDVI values < 0.2, for bare soils; the second case had NDVI values, 0.2 \( \leq \) NDVI \( \leq \) 0.5 for mixed bare soils and vegetation; and the third case had NDVI values > 0.5 for fully vegetated areas.

\[
\epsilon = \begin{cases} 
    a_i \rho_R + b_i, \\
    \epsilon_v + \epsilon_s (1 - P_V) + C, \\
    \epsilon_v + C
    \end{cases}
\]

(8)

where, \( \epsilon \) is LSE; \( \rho_R \) is the reflectance value of the red band, \( a_i \) and \( b_i \) are calculated using the empirical relationship for reflectance of red band and Moderate Resolution Imaging Spectroradiometer (MODIS) emissivity library [66].
We used the defined constant values of $\varepsilon$ from Sekertekin and Bonafoni [65] for Landsat 5 (TM) and Landsat 8 (OLI/TIRS) in given Equation (9) and Equation (10), respectively.

$$
\varepsilon = \begin{cases} 
0.979 + 0.035\rho_{R} & \text{if } NDVI < 0.2 \\
0.004P_{V} + 0.986\rho_{R} & \text{if } 0.2 \leq NDVI \geq 0.5 \\
0.99 & \text{if } NDVI > 0.5 
\end{cases} 
$$

(9)

$$
\varepsilon = \begin{cases} 
0.979 + 0.046\rho_{R} & \text{if } NDVI < 0.2 \\
0.989P_{V} + 0.977\rho_{R} & \text{if } 0.2 \leq NDVI \geq 0.5 \\
0.987 + C & \text{if } NDVI > 0.5 
\end{cases} 
$$

(10)

2.4.2. LST

To estimate LST, we used the Radiative Transfer Equation (RTE) method [65] as follows:

$$
B_{i}(T_{i}) = \frac{L_{\text{max}} - L_{\text{min}}}{Q_{CAL_{\text{max}}} - Q_{CAL_{\text{min}}}} \times (Q_{CAL} - Q_{CAL_{\text{min}}}) - L_{\text{min}} 
$$

(11)

where, $B_{i}(T_{i})$ is the spectral radiance of the top of atmosphere (TOA) (watts/(m$^2 \times$ sr $\times$ µm)), $Q_{CAL}$ is DN, $Q_{CAL_{\text{min}}}$ and $Q_{CAL_{\text{max}}}$ defines the minimum and maximum DN values of the images, respectively; $L_{\text{min}}$ and $L_{\text{max}}$ are spectral radiance of TIR band at $Q_{CAL_{\text{min}}}$ and $Q_{CAL_{\text{max}}}$ respectively: these rescaling factor values can be found in the metadata of Landsat images. Equation (11) was used for Landsat 5 (TM) for Band-6.

$$
B_{i}(T_{i}) = M_{L} \times Q_{\text{cal}} + A_{L} 
$$

(12)

where, $M_{L}$ is a multiplicative rescaling factor in the specific band from the metadata, $Q_{\text{cal}}$ is the quantized and calibrated DN values of standard product, and $A_{L}$ is additive rescaling factor in the specific band from the metadata. Equation (12) was used for Landsat 8 (OLI/TIRS) for Band-10.

We incorporated the RTE method for estimating LST as in Equation (13).

$$
B_{i}(T_{i}) = [\varepsilon B_{\lambda}(T_{s}) + (1 - \varepsilon)L_{\downarrow d}^{\tau}] \tau + L_{\uparrow u}^{\tau} 
$$

(13)

where, $B_{i}(T_{i})$ is the spectral radiance of the top of atmosphere (TOA) (watts/(m$^2 \times$ sr $\times$ µm)) for band $i$ in which have $T_{i}$ i.e., at-satellite brightness temperature; $\tau$ is atmospheric transmittance; $L_{\downarrow d}^{\tau}$ is downwelling radiance; $L_{\uparrow u}^{\tau}$ is upwelling radiance; $T_{s}$ is LST; $\varepsilon$ is emissivity of band $i$. We obtained atmospheric values such as $\tau$, $L_{\downarrow d}^{\tau}$ and $L_{\uparrow u}^{\tau}$ using an online calculator tool, called ‘Atmospheric Correction Parameter Calculator (ACPC)’ based on the given radiative transfer code of MODTRAN (http://atmcorr.gsfc.nasa.gov). $B_{\lambda}$ is Blackbody radiance at a temperature of $T_{s}$ as calculated in Equation (14) using the inversion of Equation (13).

$$
B_{\lambda}(T_{s}) = \frac{B_{i}(T_{i}) - L_{\uparrow u}^{\tau} - \tau (1 - \varepsilon)L_{\downarrow d}^{\tau}}{\varepsilon} 
$$

(14)

$T_{s}$ in Kelvin was calculated using Equation (15):

$$
T_{s} = \frac{K_{2}}{\ln \left( \frac{K_{1}}{B_{i}(T_{i}) - L_{\uparrow u}^{\tau}} + \frac{1}{\tau \varepsilon} \right)} 
$$

(15)

where, $K_{1}$ and $K_{2}$ were obtained from metadata file of respective time point images mentioned in Table 1.
2.4.3. LST Calculation (before Year 2000)

Because of the limitation of Atmospheric parameters values before the year 2000 in ACPC, we used a conventional method to extract LST values for Landsat 5 (TM) before the year 2000 only. Brightness temperature at sensor value was estimated using Equation (16) [10,13,32,41,67]:

\[
\frac{B_iT_b}{K_1} = \frac{K_2}{\ln\left(\frac{B_iT_b}{K_1} + 1\right)}
\]  

(16)

where, \(B_iT_b\) is brightness temperature (At sensor) in Kelvin; \(K_1\) and \(K_2\) are the thermal conversion constants from the metadata (Landsat 5 TM (Band 6), and Landsat 8 OLI (Band 10)) (Table 1).

Using Equation (17), the derived LST (in Kelvin (K)) through correction of emissivity was assessed from the brightness temperature [10,32,41]:

\[
T_s = \frac{B_iT_b}{1 + \frac{B_iT_b}{p} \ln(\varepsilon)}
\]  

(17)

where, \(T_s\) is temperature (At sensor) in Kelvin; \(w\) is wavelength of emitted radiance (10.8 µm for Band 10 in Landsat 8 OLI and 11.5 µm for Band 6 in Landsat 5 TM); \(p = h \times c / s\) \((1.438 \times 10^{-2} \text{mK})\), \(h\) is Planck’s constant \((6.626 \times 10^{-34} \text{Js})\), \(s\) is the Boltzmann Constant \((1.38 \times 10^{-23} \text{J/K})\), and \(c\) is the velocity of light \((2.988 \times 10^8 \text{m/s})\); \(\varepsilon\) is LSE.

Finally, we converted the \(T_s\) (LST (Kelvin)) value into Celsius (°C) using Equation (18) [10,13,32]:

\[
\text{LST (°C)} = T_s - 273.15
\]  

(18)

2.5. NDVI

An important indicator of urban climate is the Normal Difference Vegetation Index (NDVI) [10,68]. It varies between −1 and +1, where large negative (and adjacent to negative) values, positive values, and low positive values indicate water bodies, vegetation, built-up areas, or bare soils, respectively [10]. It describes information about the amount of vegetation, and its phenology and health [38]. NDVI was estimated using Equation (19) using data from the red and near infrared (NIR) bands [19,20]:

\[
\text{NDVI} = \left[\frac{\text{NIR}_{\text{Band}} - \text{Red}_{\text{Band}}}{\text{NIR}_{\text{Band}} + \text{Red}_{\text{Band}}}\right]
\]  

(19)

where, in Landsat 5 TM: NIR band denotes Band 4 (0.76–0.90 µm (wavelength)) and Red band denotes Band 3 (0.63–0.69 µm (wavelength)); while in Landsat 8 OLI: NIR band denotes Band 5 (0.85–0.88 µm (wavelength)) and Red band denotes Band 4 (0.64–0.67 µm (wavelength)).

2.6. NDBI

Another important urban climate indicator is the Normal Difference Built-up Index (NDBI) [10,68]. It varies from −1 to +1, where negative values specify water bodies and vegetation, positive value indicates the built-up area, and low positive value specifies types of bare soils [48]. It represents information about the imperviousness of the landscape [38] and was estimated using Equation (20) with NIR band and mid infrared (MIR) [19,20]:

\[
\text{NDBI} = \left[\frac{\text{MIR}_{\text{Band}} - \text{NIR}_{\text{Band}}}{\text{MIR}_{\text{Band}} + \text{NIR}_{\text{Band}}}\right]
\]  

(20)

where, in Landsat 5 TM: MIR band denotes Band 5 (1.55–1.75 µm (wavelength)) and NIR band denotes Band 4 (0.76–0.90 µm (wavelength)); while in Landsat 8 OLI: MIR band denotes Band 6 (1.57–1.65 µm (wavelength)) and NIR band denotes Band 5 (0.85–0.88 µm (wavelength)).
2.7. NDMI

Another significant urban climate indicator is the Normal Difference Moisture Index (NDMI) [39]. It also varies from $-1$ to $+1$, where a positive value specifies water bodies and vegetation and a negative value indicates built-up areas and bare soils. It signifies information about the moisture in the landscape and was estimated using Equation (21) with NIR band and mid infrared (MIR) [39]:

$$\text{NDMI} = \frac{\text{NIR Band} - \text{MIR Band}}{\text{NIR Band} + \text{MIR Band}}$$

(21)

where, in Landsat 5 TM: MIR band denotes Band 5 (1.55–1.75 $\mu$m (wavelength)) and NIR band denotes Band 4 (0.76–0.90 $\mu$m (wavelength)); while in Landsat 8 OLI: MIR band denotes Band 6 (1.57–1.65 $\mu$m (wavelength)) and NIR band denotes Band 5 (0.85–0.88 $\mu$m (wavelength)).

2.8. Analysis of Urban–Rural Gradient

The gradient approach is frequently used to evaluate spatiotemporal differences in the environment [10,48]. Here we used it to assess the spatial dynamics of mean LST, NDBI, and NDVI at 1 km intervals from the center of the city to the periphery/suburban/rural area of the city up to a maximum of 26 km (see Figure 1d for the transect of 1 km of buffers up to 26 km) [10].

2.9. Statistical Analysis (Pearson’s Correlation Coefficient)

To visualize the effects of the environmental variables (NDVI/NDBI/NDMI) on LST intensification, scatter plots were made for all time-points, (1988, 1998, 2008, and 2018) using linear regression. LST, NDBI, NDVI, and NDMI pixels were transformed into point data [10,68]. Pearson’s correlation coefficient ($r$) was used to measure the relationship among LST vs NDVI, LST vs NDBI, and LST vs NDMI, where LST was the dependent variable, and NDVI/NDBI/NDMI were the independent variables. Pearson’s ‘$r$’ was calculated through Equation (22):

$$r = \frac{\sum_{i=1}^{n} (x_i - \bar{x})(y_i - \bar{y})}{\sqrt{\sum_{i=1}^{n} (x_i - \bar{x})^2} \sqrt{\sum_{i=1}^{n} (y_i - \bar{y})^2}}$$

(22)

where, $r$ is Pearson’s correlation coefficients, $x$ represents NDVI/NDBI/NDMI measuring value of $x_i$, $y$ represents LST measuring value of $y_i$, $x_i$ and $y_i$ are single sample indexed with $i$, $\bar{x}$ and $\bar{y}$ defining the single samples indexed of $x_i$ and $y_i$, respectively.

3. Results

3.1. Accuracy Assessment of LULC Classification

LULC classification has been performed by SVM method. Accuracy assessments has been done by random sampling method by 200 sample points for each LULC classes and then error matrix has been made for each time points (i.e., 1998, 1998, 2008, and 2018). User and producer accuracy were greater than 80%, an overall accuracy was greater than 90%. Kappa coefficients were 0.90 in 1988, 0.92 in 1998, 0.94 in 2008, and 0.96 in 2018 (Appendix A Table A1).

3.2. LULC Analysis

Spatiotemporal analyses of LULC of Kathmandu Valley (Figures 2 and 3) showed that AL, VL, and IL were the major land use classes between 1988 and 2018, and that there was large increase in IL almost entirely at the expense of AL. Over this period, IL expanded by 113.44 km$^2$ (16.34% of total area) whereas AL reduced by 118.29 km$^2$ (17.03% of total area). Small changes in land were observed for OL and WB. Summaries of changes in LULC for each class are provided in Table A2, and gains and losses for each class in Table A3 and Figure A1). A small increase was observed for VL class during
1988–2018, increasing from 230.69 km$^2$ in 1988 to 236.76 km$^2$ in 2018 with a total conversion of 6.07 km$^2$
(0.87\% of total area) from other LULC classes.

Figure 2. Spatial dynamics of Land Use/Land Cover (LULC) in Kathmandu, (a) 1988; (b) 1998; (c) 2008; (d) 2018. Square box transect over all LULC maps (a–d) is to show the effects of Agriculture land on land surface temperature (LST) dynamics.

Figure 3. Information on LULC statistics of Kathmandu from 1988 to 2018 at decadal intervals (area in %).
3.3. LST Analysis

Mean LST (Table 2) increased from 23.08 °C in 1988 to 28.36 °C in 2008 and remained at that level in 2018. The spatial dynamics of changes over time in LST for the Kathmandu Valley are shown in Figure 4. The greatest transformation in LST between 1988 and 2018 was in the central, southern, and top north-eastern parts of the valley due to the massive intensification in LST of AL and IL, whereas the lowest LST was observed at the top of the northern, south-western (most bottom), and western margins of the valley because of the vast VL presence (see Figure 2). In 1988, LST values for the suburban/rural areas in Kathmandu Valley such as Katunje, Bhaktapur city, Gapali, Liwali (eastern side of the valley), Chandragiri (western side of the valley) (Figures 2 and 4), and Godavari (southern side of the valley) were in the range of 15–25 °C, whereas in the center of the valley, Kathmandu City, LST was 20–28 °C (Figure 4), giving a difference of 3–13 °C between suburban/rural areas and urban areas (Figure 4). In 1998, a difference of 5–15 °C LST was experienced across the city area, with urban areas experiencing LST values of 23–33 °C, some 18–23 °C higher than that of suburban/rural areas. In 2008, a difference of approximately 7–10 °C LST was experienced with LST in the range of 25–40 °C for urban areas, indicating a difference of 18–30 °C (Figure 4). In 2018, the city center observed a range of 25–39 °C, some 7–9 °C higher than that of the suburban/rural area (range of 18–30 °C), largely because Bhaktapur and Chandragiri had developed into dense urban areas (Figure 2).

Table 2. Maximum, minimum, and mean LST (°C) values in Kathmandu (1988–2018).

| Date       | Minimum (°C) | Maximum (°C) | Mean (°C) | Standard Deviation |
|------------|--------------|--------------|-----------|--------------------|
| 3 April 1988 | 11.88        | 35.25        | 23.08     | 3.09               |
| 15 April 1998 | 17.93        | 39.15        | 25.46     | 3.20               |
| 26 April 2008 | 17.47        | 40.30        | 28.86     | 3.31               |
| 22 April 2018 | 16.96        | 39.46        | 28.35     | 3.40               |

The Mean LST difference in different time periods (°C)

| Year       | Difference (°C) |
|------------|-----------------|
| 1988–1998  | 2.38            |
| 1998–2008  | 3.4             |
| 2008–2018  | −0.51           |
| 1988–2018  | 5.27            |
Figure 4. Spatial distribution of LST (°C) for Kathmandu in (a) 1988; (b) 1998; (c) 2008; (d) 2018. In 4 (a) the three points indicate rural/suburban areas used for the analysis of Surface Urban Heat Island (SUHI) (see Figure 2). Square box transect over all LST (°C) maps (a–d) is to show the effects of agricultural land on LST dynamics.

3.4. LULC Differences in LST

The greatest changes in LST were observed for IL (Table 3). Mean LST for IL increased from 23.8 °C in 1988 to 23.6 °C in 1998 and again to 31.0 °C in 2008 and to 30.9 °C in 2018, indicating a mean increase of 7.11 °C over the whole time period but noting this increase was not even over the time period (Table 3 and Figure 5). Similar, but slightly lower results were observed for AL, (23.7 °C, 30.4 °C, 23.8 °C, and 29.7 °C for the same time periods showing an overall increase of 6.0 °C (Table 3 and Figure 5). Changes for VL over the whole time period of 3.2 °C was much less than that for other classes, as it was 21.8 °C in 1988, 22.0 °C in 1998, 25.7 °C in 2008, 25.0 °C in 2018 (Table 3 and Figure 5). For OL and WB classes the overall change was an increase of 5.5 °C and 4.2 °C, respectively (details in Table 3).
Table 3. Mean LST for different LULC classes in Kathmandu Valley during 1988–2018.

| LULC Class         | Mean LST (°C) | The Difference of Mean LST (°C) |
|--------------------|---------------|---------------------------------|
|                    | 1988 | 1998 | 2008 | 2018 | 1988–1998 | 1998–2008 | 2008–2018 | 1988–2018 |
| Impervious Land    | 23.81 | 23.63 | 30.97 | 30.92 | −0.18 | 7.34 | −0.05 | 7.11 |
| Agriculture Land   | 23.69 | 23.80 | 30.38 | 29.67 | 0.11 | 6.58 | −0.71 | 5.98 |
| Vegetation Land    | 21.80 | 22.03 | 25.65 | 24.98 | 0.23 | 3.62 | −0.67 | 3.18 |
| Open Land          | 24.89 | 23.89 | 31.22 | 30.36 | −1 | 7.33 | −0.86 | 5.47 |
| Water Body         | 22.62 | 22.46 | 27.55 | 26.81 | −0.16 | 5.09 | −0.74 | 4.19 |

Figure 5. Spatial distribution of mean LST (°C) for different LULC classes at all time-points, (a) mean LST; (b) difference in mean LST.

Mean LST for IL was greater than that for VL by 2.1–5.9 °C and for WB by 1.2–3.4 °C at all time-points (Figure 6 and Table A4). However, mean LST was lower for IL than OL by 0.3–1.1 °C (except in 2018, where IL had a higher mean LST than OL by 0.6 °C). IL had a greater LST mean than AL of 0.1–1.3 °C at all consecutive time-points (except in 1998 see Table A4). It is evident that mean LST for VL is lower than that for all other classes of LULC on all dates. Similarly, it was evident that IL had higher mean LST than VL and WB because of the presence of the vegetation and water surface, respectively [13,69].
Jungle, Panipokhari, Samakhusi, and Amideva Budha Park. The lowest NDVI was in some central and eastern parts of the Valley, perhaps due to the large areas of AL (bare land type) and IL, respectively. The raised mean NDVI in 2018 was most likely due to the expansion of vegetation (trees and grasses) on the northern side (where Shivpuri Nagarjun National Park is located over part of Budhanilakantha, Gokarneshwor, Tokha, and Shankharapur), the western side (Nagarjun Forest Reserve), the southern side (Godavari Forest), the north-eastern side (open forest which has increased in Shankharapura and Changunarayan), and the south-western side (open forested areas increasing over part of Budhanilakantha, Gokarneshwor, Tokha, and Shankharapur), the western side (Nagarjun Forest Reserve), the southern side (Godavari Forest), the north-eastern side (open forest which has increased in Shankharapura and Changunarayan), and the south-western side (open forested areas increasing over Dakshinkali, Kirtipur, and Chandragiri). Consequently, the highest NDVI was observed in the same northern, southern, and western sides of the Valley for each date due to the presence of dense vegetation in national parks and forest reserves. However, by 2018 the area of vegetation also had increased in the central part of the valley in areas such as Gaucharan, Pashupatinath, Bhandarkhal Jungle, Panipokhari, Samakhusi, and Amideva Budha Park. The lowest NDVI was in some central and eastern parts of the Valley, perhaps due to the large areas of AL (bare land type) and IL, respectively (Figure 7).

3.5. Spatiotemporal NDVI, NDBI, and NDMI Patterns and Their Influence on LST

NDVI mean was 0.37 in 1988, increasing to 0.38 in 1998, declining to 0.36 in 2008, but increasing again to 0.43 in 2018 (Table 4). The raised mean NDVI in 2018 was most likely due to the expansion of vegetation (trees and grasses) on the northern side (where Shivpuri Nagarjun National Park is located over part of Budhanilakantha, Gokarneshwor, Tokha, and Shankharapur), the western side (Nagarjun Forest Reserve), the southern side (Godavari Forest), the north-eastern side (open forest which has increased in Shankharapura and Changunarayan), and the south-western side (open forested areas increasing over Dakshinkali, Kirtipur, and Chandragiri). Consequently, the highest NDVI was observed in the same northern, southern, and western sides of the Valley for each date due to the presence of dense vegetation in national parks and forest reserves. However, by 2018 the area of vegetation also had increased in the central part of the valley in areas such as Gaucharan, Pashupatinath, Bhandarkhal Jungle, Panipokhari, Samakhusi, and Amideva Budha Park. The lowest NDVI was in some central and eastern parts of the Valley, perhaps due to the large areas of AL (bare land type) and IL, respectively (Figure 7).

Table 4. Statistics of Land Indices and their correlation with LST.

| Date          | Statistics of NDVI and NDBI | Correlation with LST |
|---------------|-----------------------------|----------------------|
|               | Minimum | Maximum | Mean | Standard Deviation | Correlation Coefficient | Significance (p) |
| 3 April 1988  |         |         |      |                    |                       |                   |
| NDVI          | −0.073  | 0.704   | 0.365| 0.111              | −0.6665               | p < 0.001         |
| NDBI          | −0.549  | 0.313   | −0.017| 0.136              | 0.7780                | p < 0.001         |
| NDMI          | −0.313  | 0.549   | 0.017| 0.136              | −0.7780               | p < 0.001         |
| 15 April 1998 |         |         |      |                    |                       |                   |
| NDVI          | −0.074  | 0.704   | 0.367| 0.111              | −0.6527               | p < 0.001         |
| NDBI          | −0.671  | 0.416   | −0.095| 0.144              | 0.7783                | p < 0.001         |
| NDMI          | −0.416  | 0.671   | 0.095| 0.144              | −0.7783               | p < 0.001         |
| 26 April 2008 |         |         |      |                    |                       |                   |
| NDVI          | −0.042  | 0.725   | 0.356| 0.135              | −0.7678               | p < 0.001         |
| NDBI          | −0.549  | 0.318   | −0.058| 0.141              | 0.8414                | p < 0.001         |
| NDMI          | −0.318  | 0.549   | 0.058| 0.141              | −0.8414               | p < 0.001         |
| 22 April 2018 |         |         |      |                    |                       |                   |
| NDVI          | −0.025  | 0.787   | 0.430| 0.168              | −0.8045               | p < 0.001         |
| NDBI          | −0.519  | 0.378   | −0.106| 0.150              | 0.8264                | p < 0.001         |
| NDMI          | −0.378  | 0.519   | 0.106| 0.150              | −0.8264               | p < 0.001         |
The correlations between LST and NDVI were significantly negative for all time points \( p < 0.001 \), declining from \(-0.67\) in 1988, to \(-0.80\) in 2018, but increasing in 1998 to \(-0.65\) (Table 4 and Figure A2). It was very clear that reduction in vegetation cover resulted in higher LST and vice versa (Figure 4).

Mean NDBI also declined from \(-0.02\) in 1988, to 0.11 in 2018 with an increase in 2008 to \(-0.06\) (Table 4). The highest values for NDBI were in the central part in the Valley at each time-point because of densely built-up areas of Kathmandu City, Latlipur, Bhaktapur, and Madhyapur Thimi, whereas the lowest NDBI values were in some central and eastern parts of the Valley, such as Shivpuri Nagarjun National Park, Nagarjun Forest Reserve, Godavari Forest, and Bhandarkhal Jungle because of the high density VL (Figure 8).

**Figure 7.** Spatial distribution of Normal Difference Vegetation Index (NDVI) for Kathmandu during, (a) 1988; (b) 1998; (c) 2008; (d) 2018.
A significantly positive correlation was observed between LST and NDBI (Figure A3), where the coefficient of determination was greater than 0.60 at each time-point with 0.61, 0.61, 0.71, and 0.68 in 1988, 1998, 2008, and 2018 (all \( p < 0.001 \)), respectively, indicating that there is strong relationship between LST and NDBI (Table 4, Figure 4). The significant correlation between LST and NDBI (Figure A2) varied over time: 0.78, in 1988, 0.78 in 1998, and 0.84 in 2008, and 0.83 (all \( p < 0.001 \)) in 2018 (Table 4 and Figure A3). The positive correlation coefficients suggested that with increased NDBI (Figure 8), LST also increased at all time-points (Figure 4). This was particularly so in 2008 when mean LST was 3.4 °C higher than 1998 (Table 3) because the distribution of high NDBI increased due to the rapid development of IL (Figure 2 and Table A4). It is clear that less built-up areas resulted in lower LST and vice versa (Figures 2 and 4).

Mean NDMI also increased from 0.062 in 1988 to 0.101 in 2018 but an interesting drop of mean NDMI value of 0.059 was seen in 2008 (Table 4). Additionally, in 2008, mean LST increased and a
greater mean NDMI was estimated in 2018 than the mean NDMI of 2008 and this may be the reason mean LST was lower in 2018 than 2008. The lowest values for NDMI were seen over agricultural land and impervious land in the Valley at each time-point due to the densely built-up areas of Kathmandu City, Latlipur, Bhaktapur, and Madhyapur Thimi, whereas the highest NDMI values were found for vegetated areas and water bodies in areas of the Kathmandu valley such as Nagarjun Forest Reserve, Shivpuri Nagarjun National Park, Godavari Forest, and Bhandarkhal Jungle, and Bhagmati river, respectively (Figure 9).

The correlations between LST and NDMI were also significantly negative for all time points \((p < 0.001)\), declining from \(-0.78\) in 1988, to \(-0.83\) in 2018, but most decreased in 2008 to \(-0.84\) (Table 4).
and Figure A4). It was very apparent that reduction in moisture content resulted in higher LST and vice versa (Figure 4).

3.6. Analysis of Urban–Rural Gradient Pattern

Changes in LST, NDBI, NDVI, and NDMI across the city/rural gradient for at each time-point (i.e., 1988, 1998, 2008, and 2018) are shown in Figure 10 (see Figure 1d for demarcation of buffer zones). Gradient patterns are reasonably similar across the years for both NDVI and NDBI with the latter mirroring the former. However, there was much greater variation in LST between years. In 1988 and 1998 mean NDVI increased gradually with increasing distance from the city center (roughly 23–24 °C) up to around 12 km, remaining at roughly 26–27 °C until about 17 km when it gradually declined to 25 °C. In 2008 and 2018, mean NDVI was initially higher, rising more rapidly with increasing distance from 28–29 °C to around 32 °C at 17 km before declining gradually to 30–31 °C at the city periphery. Mean NDBI mirrored NDVI. In 1988 and 1998 it declined from around 23–24 °C to around 21–22 °C between 12 and 17 km then rose to around 23 °C at 22 km before declining again. In 2008 and 2018 mean NDBI at the city center was higher at 27–28 °C and declined to 25.5 °C at around 17 km before increasing to 26–27 °C at 22 km followed by a slight decline. Mean LST in the city center appeared to increase with each decadal transect. In 1988 it dropped from 25.5 °C to around 22.5 °C at 17 km and then showed a sharp increase to 26.3 °C at about 22 km before then sharply decreasing at the city boundary because of greater agricultural activities between 17–24 km (See top square box at North-Eastern side in Figure 2a–d, Figures 4a–d and 9a–d). In 1998 mean LST at the city center was higher (around 28.8 °C) and declined 24.3 °C at 15 km before rising again to 28.8 °C at 22 km before then again sharply declining. In 2008 at the center of city, mean LST was around 31.4 °C, declining to 27.8 °C at 17 km then rising sharply to 30.2 °C at 22 km and again falling to around 28 °C at 25 km due to agriculture activities, and NDMI (<0 means the lowest moisture condition) also showed lowest moisture there (Figure 9a–d). In 2018 mean LST was highest at around 32 °C within 1 km of the city center rapidly declining to 25 °C at 17 km, and as in previous decades, a further spike of 30–31 °C at around 22 km before declining again.

Consequently, it is clear that mean LST mean was dramatically greater from the city center to 15 kms compared to the surrounding area to the city periphery. Mean NDBI value was also higher in the city center to 15 km compared to the surrounding area (Figure 10), while the mean NDVI trend was the opposite, as NDVI value is lower in the area of city center up to 15 km than in the surrounding area for each decadal time-point (Figure 10). It was also very evident that mean NDVI mean was low in the areas between city center and 10 km and high between 12 km 26 km (Figure 10). Therefore, it seems clear that the highly developed VL resulted in low LST and highly developed IL resulted in higher LST and vice versa for both. Interestingly, in the 20–25 kms area, LST was again higher due to increased agriculture (Figures 2 and 10). These patterns, we believe, reflect the changes in development of built areas in the innermost parts of the city and the higher levels of vegetation towards the city periphery.
4. Discussion

We have shown that in the 30 year period between 1988 and 2018 there was a massive change in land use in the Kathmandu valley with extensive growth of IL at the cost of AL and to a lesser extent VL due to intense urbanization. At the same time, LST and NDBI have increased considerably while NDVI has reduced in core urban areas. Here we discuss the implications of our results in the context of previous studies.

4.1. Urbanization from the Perspective of LULC Transformation and Population Explosion

The capital of Nepal, Kathmandu City, is one of the fastest growing cities in South-East Asia [46, 47], located in a valley with high mountains on all sides. It is only the central part of the valley which is urbanizing fast [46, 47] and this rapid urbanization is due to religious and tourism attractions and the location of the capital administration. The rapid growth of built up areas in the form of transport networks, residential, commercial, industrial buildings, and associated parking lots have resulted in losses of some LULC classes, particularly AL, as we have shown here. This widespread transformation in LULC has led to elevated LST across a large part of the landscape [10, 38, 47, 70]. The rapid urban expansion is the result of the three-fold population increase over the last three decades in Kathmandu City (Figure A5) since it was only 0.35 million in 1988 but has increased to 1.33 million by 2018. Similar rapid population growth has occurred across the whole of the Kathmandu Valley, including the districts of Bhaktapur, Kathmandu, and Lalitpur, where the total population was 0.77 million in 1981 but had increased to 2.88 million by 2016 [49] (Figure A6). Similar levels of rapid population growth and
associated urban expansion have been observed in many other cities around the world including Chennai in India [71]; Lucknow in India [3,72]; Agra in India [73]; Xuzhou in China [74]; Baguio in the Philippines [75]; Kandy City in Sri Lanka [48]; Tokyo in Japan [11]; Tehran in Iran [10]; Istanbul in Turkey [76]; Sobotka in Poland [77]; Santiago de Chile [78]; Mekelle in Ethiopia [79]; Bucharest city in Romania, Budapest city in Hungary, Prague city in Czech Republic, Sofia city in Bulgaria, and Warsaw city in Poland [80]; and São José dos Campos in Brazil [81].

We have shown that IL in the last three decades, especially most recently in 2008 and 2018, has swiftly extended towards the eastern side (e.g., Madhyapur Thimi, Katunje, Bhaktapur, Liwali, Gapali, Jhaukhel, and Duwakot), the western side (e.g., Chandragiri, Kirtipur, Sitapalia, Gairi Gaun, Boharatar, and Ramkot), the northern side (e.g., Kapan, Dharampur, Bhangal, Thapa Gaun, Mahankal, Gongabu, Paiyatar, and Hiledol), and the southern side (e.g., Patan, Bhaisepati, Kusunti, Mitra Tole, Sunakothi, Dhapakhel, Thaiba, Thecho, and Bullu) (Figure 2). Such growth is predicted to continue as the United Nations has projected that Kathmandu City population will grow to 2.2 million by 2035 [4], roughly doubling the 2018 population (Figure A5), whereas NPCS has predicted the population for the whole of the Kathmandu Valley to increase to 3.85 million by 2031 (Figure A6).

4.2. Phenomena of SUHI and Sustainable Planning

Our results suggest that the rapid expansion of built-up areas and their influence in increasing SUHI is occurring at a large scale in Kathmandu Valley. This is largely due to the conversion of AL and, to a lesser extent, VL into IL in the form of transport networks, industrial, commercial, residential, parking lots, and other paved surfaces. We found that both VL and WB had lower mean LST than IL, AL, and OL at all studied sequential time-points. This contrasts with the results of other studies which indicated that IL had higher LST than others classes of LULC especially VL and WB, in tropical montane cities, such as, Tehran in Iran [10], Kandy in Sri Lanka [48], and Baguio in Philippines [75]. We noted that LST was higher in 2008 and we recognize the limitation in our study that in selecting single dates for our Landsat images we cannot account for daily variations in LST that are naturally likely to occur. However, overall mean LST increased over the whole 40-year time period. However, there could be other factors (like, elevation) influencing LST intensification which could be examined in the future.

We found that LST intensified throughout the central city area of the Kathmandu Valley as the IL rapidly increased. In many western cities, attempts to reduce the effects of SUHI have included strategies such as introducing more street trees or growing plants on roofs, so-called green roofs, as well as developing materials that cool roofs and cool pavements, and the use of light materials [82]. Other cooling mechanisms included improving wind flow, by carefully designing the size, shape, and orientation of buildings [83]. New cities such as Kathmandu which have grown with little planning need to recognize the value of such greening designs with the opportunity to enhance environmental sustainability in the city [10,18,41,68].

We also found that the mean LST was always higher in the city center than its periphery and that in general LST has risen over the 40-year period. In between 12–18 km from the city center, LST level increased due to increased NDBI because of greater population pressure resulting in urban expansion. Kathmandu Valley has also observed an increase in SUHI from the city center towards the city periphery. The reason behind this intensification was due to dense urban development (especially in the areas of Kaldhara, Chhetrapati, Sanepa, Kalimati, Siddhitol, Bhimsengola, Sinamangal, Kumariapat, Ekantakuna, and Narephate) and at the periphery of the city with the loss of green spaces (especially in the areas of Budhanilkantha, Tarakeswar, Sitapalia, Charchare, Bhatkapi, Suyel Gaun, Dadhikot, Chhaling, and Taudol).
4.3. Urban Sustainability Implication

Typically, in the relatively unplanned growth of built-up areas micro- to macro- level features develop to reflect economic and social needs \cite{10,32,45,84}. Greater employment opportunities and higher economic development are important reasons why people are drawn to urban areas but at the same time the resulting reduction in AL and VL leads to intensification of SUHI and consequent harm to the local environment and the urban population \cite{10,85,86}.

With so few open areas available for development in the city center we found that Kathmandu valley has had higher IL development at the periphery than in the city center between 1988 and 2018. As a result, from 6–8 kms, IL expanded largely at the cost of AL. This partly explains some of the observed changes in LULC statistics (Figure 2 and Table A2).

Our results suggest that SUHI has intensified at all-time points resulting from the increase in IL at the cost of VL and AL which is likely to have severe consequences for the local environment. Sustainable land management practices can decrease the negative effects of stressors like climate change \cite{7}. Therefore, to minimize some of the severe effects of SUHI novel greening strategies need to be developed, as discussed above, as well as reducing runoff and enhancing availability of freshwater through the creation of ponds/lakes, and use of rain water harvesting. Such structures can enhance the resilience of local ecosystems \cite{9,10,18}. Increased public awareness regarding the severe effects of SUHI phenomena may result in pressure at the local and national level on public or private authorities to reduce the SUHI phenomena through adoption of remedies to minimize its effects as well as improving future sustainable planning for new urban areas \cite{10}.

5. Conclusions

We have shown a significant negative and positive relationship between LST and NDVI, and LST and NDBI, respectively. We interpret this to show that this indicates that vegetation had a very significant role in decreasing LST by nearly 0.82–5.94 °C compared to other LULC classes over the last 30 years in the study area. At the same time, we are concerned about the increase in mean LST of 0.12–5.94 °C for built-up areas. Our results show that there was a sharp decrease of mean LST with NDBI from city center to 15 kms with NDVI showing the opposite pattern sharply increasing from the city center to 15 kms, clearly demonstrating the creation of SUHI. Further, at around 20–25 kms from the city center, the mean LST again rose due to the intensive agriculture there as we have found mean NDVI and as well as mean NDMI also declined there, which reveals agriculture activities’ effects on higher LST distribution. It was very apparent that because of explosive increase in IL there was a resulting loss in AL and VL. One noteworthy observation is that OL had greater mean LST than all other classes of LULC at all consecutive time points. We conclude that it is essential to measure thermal state for cities over time to depict LULC and SUHI creation because it can explain the consequences of historical changes in the city’s landscape. The patterns of LULC and LST observed in this study we hope will be useful for future urban planning and policy making in the Kathmandu Valley.

The enormous changes in urbanization resulting from rapidly increasing population growth in the Kathmandu valley and their consequences for LULC and LST change are of great concern particularly as population growth is predicted to continue. To avoid severe consequences of SUHI, strong planning policies and actions need to be taken to protect current urban spaces, reduce vegetation depletion, and open space reduction. At the same time, efforts are needed to reduce SUHI by improving building design using green city technologies. At the same time, reduction in runoff and improved rainwater harvesting will be essential at both local and large scales through the participation of individuals, private organizations, and local to national government.

Supplementary Materials: The following are available online at http://www.mdpi.com/2220-9964/9/12/726/s1. NASA’s POWER project weather information over selected sample location (Latitude: 27.7017 and Longitude: 85.319) of selected time points between 1988 and 2018 for this study area, Kathmandu Valley area available in excel sheet in the given this link.
Author Contributions: The topic has Conceptualized by Md. Omar Sarif and Bhagawat Rimal.; Methodology, Md. Omar Sarif and Bhagawat Rimal (LULC); Validation, Md. Omar Sarif and Bhagawat Rimal; Formal Analysis, Md. Omar Sarif and Bhagawat Rimal; Investigation, Md. Omar Sarif; Resources, Md. Omar Sarif and Bhagawat Rimal; Data Curation, Md. Omar Sarif and Bhagawat Rimal; Writing-Original Draft Preparation, Md. Omar Sarif and Bhagawat Rimal; Writing-Review & Editing, Md. Omar Sarif, Bhagawat Rimal, Nigel E. Stork; Visualization, Md. Omar Sarif and Bhagawat Rimal; All authors have read and agreed to the published version of the manuscript.

Funding: This research received no external funding.

Acknowledgments: The authors are grateful to the United States Geological Survey (USGS) for making Landsat datasets freely available. Authors are also grateful to NASA Langley Research Center (LaRC) POWER Project funded through the NASA Earth Science/Applied Science Program for weather information using its online portal (https://power.larc.nasa.gov/data-access-viewer/). We would like to thank for anonymous reviewers for their invaluable comments and suggestion.

Conflicts of Interest: Authors have no conflict of interest.

Appendix A

Figure A1. Gains and losses of LULC classes in Kathmandu at decadal intervals from 1988 to 2018, (a) area (in Km²); (b) percentage area.

Figure A2. Statistical Scatter Plots between LST and NDVI for (a) 1988; (b) 1998; (c) 2008; (d) 2018.
Figure A3. Scatter Plots between LST and NDBI for (a) 1988; (b) 1998; (c) 2008; (d) 2018.

Figure A4. Scatter Plots between LST and NDMI for (a) 1988; (b) 1998; (c) 2008; (d) 2018.
Figure A5. Population of Kathmandu City (1950–2035), Source: World Urbanization Prospects: The 2018 Revision by United Nation in 2018.

Figure A6. Population of Kathmandu Valley (includes Bhaktapur, Kathmandu, and Lalitpur) in Nepal during 1981–2031 (Source: NPHC 2011, Central Bureau of Statistics, National Planning Commission Secretarial, Government of Nepal).

Table A1. Accuracy assessment report of Land Use/Land Cover classification in Kathmandu (1988–2018).

| LULC Class | 1988 | 1998 | 2008 | 2018 |
|------------|------|------|------|------|
| User Accuracy (%) |      |      |      |      |
| IL         | 88.0 | 89.0 | 91.0 | 96.0 |
| AL         | 95.5 | 96.0 | 96.5 | 96.0 |
| VL         | 84.5 | 90.0 | 95.5 | 96.5 |
| OL         | 93.5 | 94.0 | 95.0 | 96.0 |
| WB         | 97.5 | 98.0 | 96.5 | 98.5 |
| Producer Accuracy (%) |      |      |      |      |
| IL         | 92.6 | 87.3 | 94.8 | 93.7 |
| AL         | 83.8 | 86.9 | 91.0 | 93.7 |
| VL         | 90.0 | 95.7 | 96.4 | 97.8 |
| OL         | 99.4 | 99.5 | 99.5 | 100  |
| WB         | 94.7 | 97.0 | 98.0 | 98.5 |
| Overall Accuracy (%) | 95.2 | 93.4 | 95.2 | 96.6 |
| Kappa Coefficient | 0.897 | 0.917 | 0.940 | 0.958 |
Table A2. The summary Land Use/Land Cover statistics of Kathmandu at 10-years interval from 1988 to 2018.

| Year | IL  | AL  | VL  | OL  | WB  | Total |
|------|-----|-----|-----|-----|-----|-------|
| Km²  | %   | Km² | %   | Km² | %   | Km²   |
| 1988 | 37.98 | 418.32 | 230.69 | 1.97 | 5.31 | 694.27 |
| 1998 | 5.47  | 60.25  | 33.23  | 0.28 | 0.76 | 100   |
| 2008 | 8.49  | 56.38  | 33.97  | 0.5  | 0.66 | 100   |
| 2018 | 13.75 | 51.37  | 33.91  | 0.48 | 0.5  | 100   |

Table A3. Summary Land Use/Land Cover change statistics for Kathmandu Valley.

| Year       | IL  | AL  | VL  | OL  | WB  | Total |
|------------|-----|-----|-----|-----|-----|-------|
| Km²        | %   | Km² | %   | Km² | %   | Km²   |
| 1988–1998  | 20.99 | −26.86 | 5.16  | 1.48 | −0.76 |       |
| 1998–2008  | 3.02  | −3.87 | 0.74  | 0.22 | −0.1  |       |
| 2008–2018  | 5.26  | −5.01 | −0.06 | −0.02 | −0.16 |       |
| 1988–2018  | 8.06  | −8.15 | 0.19  | −0.04 | −0.07 |       |

Table A4. Comparisons of mean LST between Land Use/Land Cover classes from 1988 to 2018.

| Comparison of LULC Class | LULC Class (Cross Cover Comparison) | Mean LST Magnitude (°C) |
|--------------------------|------------------------------------|-------------------------|
|                          | 1988 | 1998 | 2008 | 2018 |
| IL vs. Other Class       |      |      |      |      |
| IL-VL                    | 2.01 | 1.6  | 5.32 | 5.94 |
| IL-WB                    | 1.19 | 1.17 | 3.42 | 4.11 |
| IL-AL                    | 0.12 | −0.17| 0.59 | 1.25 |
| IL-OL                    | −1.08| −0.26| −0.25| 0.56 |
| VL vs. Other Class       |      |      |      |      |
| VL-IL                    | −2.01| −1.6 | −5.32| −5.94|
| VL-WB                    | −1.89| −1.77| −4.73| −4.69|
| VL-AL                    | −3.09| −1.86| −5.57| −5.38|
| VL-OL                    | −0.82| −0.43| −1.9 | −1.83|
| WB vs. Other Class       |      |      |      |      |
| WB-IL                    | −1.19| −1.17| −3.42| −4.11|
| WB-WB                    | −1.07| −1.34| −2.83| −2.86|
| WB-AL                    | 0.82 | 0.43 | 1.9  | 1.83 |
| WB-OL                    | −2.27| −1.43| −3.67| −1.55|

References
1. Feng, Y.; Du, S.; Myint, S.W.; Shu, M. Do urban functional zones affect land surface temperature differently? A case study of Beijing, China. Remote Sens. 2019, 11, 1802. [CrossRef]
2. Ranagalage, M.; Wang, R.; Gunarathna, M.H.J.P.; Dissanayake, D.; Murayama, Y.; Simwanda, M. Spatial Forecasting of the Landscape in Rapidly Urbanizing Hill Stations of South Asia: In Proceedings of the A Case Study of Nuwara Eliya, Sri Lanka (1996–2037). Remote Sens. 2019, 11, 1743. [CrossRef]
3. Shukla, A.; Jain, K. Modeling Urban Growth Trajectories and Spatiotemporal Pattern: A Case Study of Lucknow City, India. J. Indian Soc. Remote Sens. 2019, 47, 139–152. [CrossRef]
4. UN Department of Economic and Social Affairs. *Population Division. World Urbanization Prospects: The 2018 Revision.* Online Edition. 2018. Available online: https://population.un.org/wup/ (accessed on 15 November 2020).

5. Wang, C.; Myint, S.W.; Wang, Z.; Song, J. Spatio-Temporal Modeling of the Urban Heat Island in the Phoenix Metropolitan Area: Land Use Change Implications. *Remote Sens.* 2016, 8, 185. [CrossRef]

6. Keeratikasikorn, C.; Bonafoni, S. Urban heat island analysis over the land use zoning plan of Bangkok by means of Landsat 8 imagery. *Remote Sens.* 2018, 10, 440. [CrossRef]

7. IPCC Climate Change and Land. *An IPCC Special Report on climate change, desertification, land degradation, sustainable land management, food security, and greenhouse gas fluxes in terrestrial ecosystems. In Summary for Policymakers.* 2019. Available online: https://www.ipcc.ch/site/assets/uploads/sites/4/2020/02/SPM_Updated-Jan20.pdf (accessed on 15 November 2020).

8. Weng, Q.; Lu, D.; Schubring, J. Estimation of land surface temperature-vegetation abundance relationship for urban heat island studies. *Remote Sens. Environ.* 2004, 89, 467–483. [CrossRef]

9. Estoque, R.C.; Murayama, Y.; Myint, S.W. Effects of landscape composition and pattern on land surface temperature: An urban heat island study in the megacities of Southeast Asia. *Sci. Total Environ.* 2017, 577, 349–359. [CrossRef]

10. Rousta, I.; Sarif, M.O.; Gupta, R.D.; Olafsson, H.; Ranagalage, M.; Murayama, Y.; Zhang, H.; Mushore, T.D. Spatiotemporal Analysis of Land Use/Land Cover and Its Effects on Surface Urban Heat Island Using Landsat Data: A Case Study of Metropolitan City Tehran (1988–2018). *Sustainability* 2018, 10, 4433. [CrossRef]

11. Wang, R.; Derdouri, A.; Murayama, Y. Spatiotemporal simulation of future land use/cover change scenarios in the Tokyo metropolitan area. *Sustainability* 2018, 10, 2056. [CrossRef]

12. Rimal, B.; Keshtkar, H.; Sharma, R.; Stork, N.; Rijal, S.; Kunwar, R. Simulating urban expansion in a rapidly changing landscape in eastern Tarai, Nepal. *Environ. Monit. Assess.* 2019, 191, 255. [CrossRef]

13. Bokaeia, M.; Zarkesh, M.K.; Arasteh, P.D.; Hosseini, A. Assessment of urban heat island based on the relationship between land surface temperature and land use/cover in Tehran. *Sustain. Cities Soc.* 2016, 23, 94–104. [CrossRef]

14. Son, N.T.; Chen, C.F.; Chen, C.R.; Thanh, B.X.; Vuong, T.H. Assessment of urbanization and urban heat islands in Ho Chi Minh city, Vietnam using Landsat data. *Sustain. Cities Soc.* 2017, 30, 150–161. [CrossRef]

15. Babazadeh, M.; Kumar, P. Estimation of the urban heat island in local climate change and vulnerability assessment for air quality in Delhi. *Eur. Sci. J.* 2015, 1, 55–65. [CrossRef]

16. Stewart, I.D.; Oke, T.R. Local climate zones for urban temperature studies. *Bull. Am. Meteorol. Soc.* 2012, 93, 1879–1900. [CrossRef]

17. Joshi, R.; Raval, H.; Pathak, M.; Prajapati, S.; Patel, A.; Singh, V.; Kalubarme, M.H. Urban heat island characterization and isotherm mapping using geo-informatics technology in Ahmedabad city, Gujarat state, India. *Int. J. Geosci.* 2015, 6, 274–285. [CrossRef]

18. Rosa, A.; De Oliveira, P.S.; Gomes, A.; Gleriani, J.M.; Gonçalves, W.; Moreira, G.L.; Silva, F.G.; Ricardo, E.; Branco, F.; Moura, M.M.; et al. Spatial and temporal distribution of urban heat islands. *Sci. Total Environ.* 2017, 605–606, 946–956.

19. Avdan, U.; Jovanovska, G. Algorithm for automated mapping of land surface temperature using Landsat 8 satellite data. *J. Sens.* 2016, 2016, 1–8. [CrossRef]

20. Li, X.; Zhou, Y.; Asrar, G.R.; Imhoff, M.; Li, X. The surface urban heat island response to urban expansion: A panel analysis for the conterminous United States. *Sci. Total Environ.* 2017, 605–606, 426–435. [CrossRef]

21. Zhang, X.; Estoque, R.C.; Murayama, Y. An urban heat island study in Nanchang City, China based on land surface temperature and social-ecological variables. *Sustain. Cities Soc.* 2017, 32, 557–568. [CrossRef]

22. Gagliano, A.; Detommaso, M.; Nocera, F.; Evola, G. A multi-criteria methodology for comparing the energy and environmental behavior of cool, green and traditional roofs. *Build. Environ.* 2015, 90, 71–81. [CrossRef]

23. Chaudhuri, A.S.; Singh, P.; Rai, S.C. Modelling LULC change dynamics and its impact on environment and water security: Geospatial technology based assessment. *Ecol. Environ. Conserv.* 2018, 24, 300–306.

24. Yao, R.; Wang, L.; Huang, X.; Guo, X.; Niu, Z.; Liu, H. Investigation of urbanization effects on land surface phenology in northeast China during 2001–2015. *Remote Sens.* 2017, 9, 66. [CrossRef]

25. Dewan, A.M.; Yamaguchi, Y. Land use and land cover change in Greater Dhaka, Bangladesh: Using remote sensing to promote sustainable urbanization. *Appl. Geogr.* 2020, 29, 390–401. [CrossRef]
26. Ramachandra, T.V.; Kumar, U. Land surface temperature with land cover dynamics: Multi-resolution, ratio-temporal data analysis of Greater Bangalore. *Int. J. Geoinform.* 2009, 5, 43–53.

27. Pandey, P.; Kumar, D.; Prakash, A.; Masih, J.; Singh, M.; Kumar, S.; Jain, V.K.; Kumar, K. Science of the Total Environment: A study of urban heat island and its association with particulate matter during winter months over Delhi. *Sci. Total Environ.* 2012, 414, 494–507. [CrossRef]

28. Ku, D.N.; Sandeep, N.; Jyothi, S.; Madhu, T. Significant changes on land use/land cover by using remote sensing and GIS analysis-review. *Int. J. Eng. Sci. Comput.* 2017, 7, 5433–5435.

29. Gould, W.A.; Gonz, O.M.R. Land development, land use, and urban sprawl in Puerto Rico integrating remote sensing and population census data. *Landsc. Urban Plan.* 2007, 79, 288–297.

30. Kuang, W.; Liu, Y.; Dou, Y.; Chi, W.; Chen, G. What are hot and what are not in an urban landscape: Quantifying and explaining the land surface temperature pattern in Beijing, China. *Landsc. Ecol.* 2015, 30, 357–373. [CrossRef]

31. Zhang, Y.; Su, Z.; Li, G.; Zhuo, Y.; Xu, Z. Spatial-temporal evolution of sustainable urbanization development: A perspective of the coupling development based on population, industry, and built up land spatial agglomeration. *Sustainability* 2018, 10, 1766. [CrossRef]

32. Sultana, S.; Satyanarayana, A.N.V. Urban heat island intensity during winter over metropolitan cities of India using remote-sensing techniques: Impact of urbanization. *Int. J. Remote Sens.* 2018, 39, 6692–6730. [CrossRef]

33. Kumar, R.; Mishra, V.; Buzan, J.; Kumar, R.; Shindell, D.; Huber, M. Dominant control of agriculture and irrigation on urban heat island in India. *Sci. Rep.* 2017, 7, 14054. [CrossRef] [PubMed]

34. Kumar, S.; Panwar, M. Urban heat island footprint mapping of Delhi using remote sensing. *Int. J. Emerg. Technol.* 2017, 8, 80–83.

35. Agarwal, R.; Sharma, U.; Taxak, A. Remote sensing based assessment of urban heat island phenomenon in Nagpur metropolitan area. *Int. J. Inf. Comput. Technol.* 2014, 4, 1069–1074.

36. Mukherjee, S.; Joshi, P.K.; Garg, R.D. Analysis of urban built-up areas and surface urban heat island using downscaled MODIS derived land surface temperature data. *Geocarto Int.* 2017, 32, 900–918. [CrossRef]

37. Gunaalan, K.; Ranagalage, M.; Gunarathna, M.H.J.P.; Kumari, M.K.N.; Vithanage, M.; Srivaratharasan, T.; Saravanan, S.; Warnasuriya, T.W.S. Application of geospatial techniques for groundwater quality and availability assessment: A case study in Jaffna Peninsula, Sri Lanka. *ISPRS Int. J. Geo-Inform.* 2018, 7, 20. [CrossRef]

38. Pal, S.; Ziaul, S. Detection of land use and land cover change and land surface temperature in English Bazar urban centre. *Egypt. J. Remote Sens. Sp. Sci.* 2017, 20, 125–145. [CrossRef]

39. Renard, F.; Alonso, L.; Fitts, Y.; Hadjiosif, A.; Comby, J. Evaluation of the Effect of Urban Redevelopment on Surface Urban Heat Islands. *Remote Sens.* 2019, 11, 1–31. [CrossRef]

40. Thapa, R.B.; Murayama, Y. Drivers of urban growth in the Kathmandu valley, Nepal: Examining the efficacy of the analytic hierarchy process. *Appl. Geogr.* 2010, 30, 70–83. [CrossRef]

41. Estoque, R.C.; Pontius, R.G.; Murayama, Y.; Hou, H.; Thapa, R.B.; Lasco, R.D.; Villar, M.A. Simultaneous comparison and assessment of eight remotely sensed maps of Philippine forests. *Int. J. Appl. Earth Obs. Geoinf.* 2018, 67, 123–134. [CrossRef]

42. Toffin, G. Urban fringes: Squatter and slum settlements in the Kathmandu Valley (Nepal). *Contrib. Nepal. Stud.* 2010, 37, 151–168.

43. Thapa, R.B.; Murayama, Y. Examining Spatiotemporal Urbanization Patterns in Kathmandu Valley, Nepal: Remote Sensing and Spatial Metrics Approaches. *Remote Sens.* 2009, 1, 534–556. [CrossRef]

44. Chitrakar, R.M.; Baker, D.C.; Guaralda, M. Urban growth and development of contemporary neighbourhood public space in Kathmandu Valley, Nepal. *Habitat Int.* 2016, 53, 30–38. [CrossRef]

45. Rimal, B.; Zhang, L.; Keshtkar, H.; Wang, N.; Lin, Y. Monitoring and modeling of spatiotemporal urban expansion and Land-Use/Land-Cover change using integrated Markov Chain Cellular Automata Model. *ISPRS Int. J. Geo-Inf.* 2017, 6, 288. [CrossRef]

46. Rimal, B.; Zhang, L.; Keshtkar, H.; Barry, N.H.; Rijal, S.; Zhang, P. Land Use/Land Cover Dynamics and Modeling of Urban Land Expansion by the Integration of Cellular Automata and Markov Chain. *ISPRS Int. J. Geo-Inf.* 2018, 7, 154. [CrossRef]

47. Ishhtiaque, A.; Shrestha, M.; Chhetri, N. Rapid urban growth in the Kathmandu valley, Nepal: Monitoring land use land cover dynamics of a himalayan city with Landsat imageries. *Environments* 2017, 4, 72. [CrossRef]
62. Nimish, G.; Bharath, H.A.; Lalitha, A. Exploring temperature indices by deriving relationship between land surface temperature and urban landscape. Remote Sensing Applications: Society and Environment. 2018, 7, 341. [CrossRef]

49. NPS National Population and Housing Census 2011 (Population Projection 2011–2031); Central Bureau of Statistics, Government of Nepal: Kathmandu, Nepal, 2014; Volume 8.

48. Ranagalage, M.; Dmslb, D.; Murayama, Y.; Zhang, X.; Estoque, R.C.; Enc, P.; Morimoto, T. Quantifying urban heat island formation in the world heritage tropical mountain city of Sri Lanka. ISPRS Int. J. Geo-Inf. 2020, 9, 726.
69. Roshan, G.; Rousta, I.; Ramesh, M. Studying the effects of urban sprawl of metropolis on tourism-climate index oscillation: A case study of Tehran city. J. Geogr. Reg. Plan. 2009, 2, 310–321.
70. Li, D.; Sun, T.; Liu, M.; Yang, L.; Wang, L.; Gao, Z. Contrasting responses of urban and rural surface energy budgets to heat waves explain synergies between urban heat islands and heat waves. Environ. Res. Lett. 2015, 10, 054009. [CrossRef]
71. Padmanaban, R.; Bhowmik, A.K.; Cabral, P.; Zamyatin, A.; Almegdadi, O.; Wang, S. Modelling urban sprawl using remotely sensed data: A case study of Chennai city, Tamilnadu. Entropy 2017, 19, 163. [CrossRef]
72. Singh, P.; Kikon, N.; Maurya, G.; Rathore, A.; Sarif, M.O.; Gupta, R.D. The Effects of Land Indices on Thermal State in Surface Urban Heat Island Formation: A Case Study on Agra City in India Using Remote Sensing Data (1992–2019). Earth Syst. Environ. 2020. [CrossRef]
73. Xi, Y.; Thinh, N.X.; Li, C. Spatio-Temporal Variation Analysis of Landscape Pattern Response to Land Use Change from 1985 to 2015 in Xuzhou City, China. Sustainability 2018, 10, 4287. [CrossRef]
74. Estoque, R.C.; Murayama, Y. Monitoring surface urban heat island formation in a tropical mountain city using Landsat data (1987–2015). ISPRS J. Photogramm. Remote Sens. 2017, 133, 18–29. [CrossRef]
75. Fenta, A.A.; Yasuda, H.; Haregeweyn, N.; Belay, S.; Hadush, Z.; Gebremedhin, M.A. The dynamics of urban expansion and land use/land cover changes using remote sensing and spatial metrics: The case of Mekelle City of northern Ethiopia changes using remote sensing and spatial metrics: The case of Mekelle City of northern Ethiopia. Int. J. Remote Sens. 2017, 38, 4107–4129. [CrossRef]
76. Garcia-Ayllon, S. Urban transformations as indicators of economic change in post-communist Eastern Europe: Territorial diagnosis through five case studies. Habitat Int. 2018, 71, 29–37. [CrossRef]
77. Ogashawara, I.; da Bastos, V.S.B. A Quantitative Approach for Analyzing the Relationship between Urban Heat Islands and Land Cover. Remote Sens. 2012, 4, 3596–3618. [CrossRef]
78. Mohajerani, I.; Bakaric, J.; Jeffrey-Bailey, T. The urban heat island effect, its causes, and mitigation, with reference to the thermal properties of asphalt concrete. J. Environ. Manag. 2017, 197, 522–538. [CrossRef]
79. Sen, S.; Roesler, J.; Ruddell, B.; Middel, A. Cool Pavement Strategies for Urban Heat Island Mitigation in Suburban Phoenix, Arizona. Sustainability 2019, 11, 4452. [CrossRef]
80. Rimal, B.; Sloan, S.; Keshtkar, H.; Sharma, R.; Rijal, S.; Shrestha, U.B. Patterns of historical and future urban expansion in Nepal. Remote Sens. 2020, 12, 628. [CrossRef]
81. Ranagalage, M.; Estoque, R.C.; Handayani, H.H.; Zhang, X.; Morimoto, T.; Tadono, T.; Murayama, Y. Relation between urban volume and land surface temperature: A comparative study of planned and traditional cities in Japan. Sustainability 2018, 10, 2366. [CrossRef]
82. Dissanayake, D.; Morimoto, T.; Murayama, Y.; Ranagalage, M. Impact of Landscape Structure on the Variation of Land Surface Temperature in Sub-Saharan Region: A Case Study of Addis Ababa using Landsat Data. Sustainability 2019, 11, 2257. [CrossRef]

Publisher’s Note: MDPI stays neutral with regard to jurisdictional claims in published maps and institutional affiliations.

© 2020 by the authors. Licensee MDPI, Basel, Switzerland. This article is an open access article distributed under the terms and conditions of the Creative Commons Attribution (CC BY) license (http://creativecommons.org/licenses/by/4.0/).