Twisted tale of the tiger: the case of inappropriate data and deficient science

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Publications in peer reviewed journals are often looked upon as tenets on which future scientific thought is built. Published information is not always flawless and errors in published research should be expediently reported, preferably by a peer review process. We review a recent publication by Gopalaswamy et al (2015; doi:10.1111/2041-210X.12351) that challenges the use of “double sampling” in large scale animal surveys. Double sampling is often resorted to as an established economical and practical approach for large scale surveys since it calibrates abundance indices against absolute abundance, thereby potentially addressing the statistical shortfalls of indices. Empirical data used by Gopalaswamy et al. (2015) to test their theoretical model, relate to tiger sign and tiger abundance referred to as an Index Calibration experiment (IC-Karanth). These data on tiger abundance and signs should be paired in time and space to qualify as a calibration experiment for double sampling, but original data of IC-Karanth show lags of (up to) several years. Further, data points used in the paper do not match the original sources. We show that by use of inappropriate and incorrect data collected through a faulty experimental design, poor parameterization of their theoretical model, and selectively-picked estimates from literature on detection probability, the inferences of this paper are highly questionable. We highlight how the results of Gopalaswamy et al. were further distorted in popular media. If left unaddressed, Gopalaswamy et al. paper could have serious implications on statistical design of large-scale animal surveys by propagating unreliable inferences.
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Abstract:
Publications in peer reviewed journals are often looked upon as tenets on which future scientific thought is built. Published information is not always flawless and errors in published research should be expediently reported, preferably by a peer review process. We review a recent publication by Gopalaswamy et al (2015; doi:10.1111/2041-210X.12351) that challenges the use of “double sampling” in large scale animal surveys. Double sampling is often resorted to as an established economical and practical approach for large scale surveys since it calibrates abundance indices against absolute abundance, thereby potentially addressing the statistical shortfalls of indices. Empirical data used by Gopalaswamy et al. (2015) to test their theoretical model, relate to tiger sign and tiger abundance referred to as an Index Calibration experiment (IC-Karanth). These data on tiger abundance and signs should be paired in time and space to qualify as a calibration experiment for double sampling, but original data of IC-Karanth show lags of (up to) several years. Further, data points used in the paper do not match the original sources. We show that by use of inappropriate and incorrect data collected through a faulty experimental design, poor parameterization of their theoretical model, and selectively-picked estimates from literature on detection probability, the inferences of this paper are highly questionable. We highlight how the results of Gopalaswamy et al. were further distorted in popular media. If left unaddressed, Gopalaswamy et al. paper could have serious implications on statistical design of large-scale animal surveys by propagating unreliable inferences.
Scientific method operates by testing competing hypothesis or by choosing between alternate models that best explain observed data. Hypothesis and models that survive repeated testing by careful experimentation are published through rigorous scrutiny by a peer review process, these subsequently become scientific theory (Gauch 2012). An incorrect experimental design, inappropriate data collection protocol, and selective data, used for analysis from telemetered Florida panthers (*Puma concolor*) (Gross 2005) resulted in a peer reviewed publication of habitat use and preference (Maehr & Cox 1995) in Conservation Biology. The results were subsequently used for land use planning and policy (Maehr & Deason 2002) which resulted in the best panther habitat being lost to developmental projects (Gross 2005). In the ideal world, response to deficiencies in science is best made through a peer review process, since scientists understand the intricacies of the scientific method probably more than others (Parsons & Wright 2015).

In a recent paper "An examination of index-calibration experiments: counting tigers at macroecological scales" published in the journal Methods of Ecology and Evolution, Gopalaswamy et al (2015a) supposedly demonstrate that as part of their long-term, large-scale data on tiger abundance and index (IC-Karanth) they did not find any relationship between tiger abundance and scat index. They conclude that attempting to use double sampling (Cochran 1977; Eberhardt & Simmons 1987; Pollock et al. 2002) to establish relationships between any index of abundance and actual abundance is a futile effort. In particular, they claim that the relationship between tiger sign index and tiger abundance published by Jhala et al. (2011a) to be improbable since they could not reproduce it by their data or theoretical model. We review Gopalaswamy et al (2015a) to show that by the use of a) wrong ecological parameters for their theoretical model, b) selectively picked references from literature, c) inappropriate and incorrect data, and d) data not collected in an experimental setup, the inferences drawn by their paper are questionable.

**a) Use of inadequate Ecological Parameters:**

The basic premise for index calibration by double sampling is that animal sign intensity or count data should reflect underlying animal abundance. Often due to logistic and economic constraints, large scale estimates of abundance are not possible through statistically rigorous methods that
explicitly estimate and correct for detection (e.g. capture-mark-recapture or DISTANCE sampling). Double sampling approach as described initially by Cochran (1977) and applied to wildlife surveys by Eberhardt & Simmons (1987), allows us to address this limitation by measuring a relatively easy and economically less expensive, but potentially biased index of abundance across all sampling units, while simultaneously estimating detection corrected abundance from within a subset of these sampling units (Conroy & Carroll 2009, Williams et al. 2002). Subsequently, the potentially biased index is calibrated against the unbiased estimate of abundance or actual abundance using a ratio or regression approach (Skalski et al. 2005). Pollock et al. (2002) recommend double sampling as a sensible large-scale survey design for most species.

To prove their point of view, Gopalaswamy et al (2015a) use detection probability (p) estimates from tiger occupancy studies as a surrogate for detection probability of tiger scat for parameterizing their theoretical model. This p is the probability of finding (or not finding) tiger sign on a single survey in an area occupied by tigers. Gopalaswamy et. al. (2015) confuse p of occupancy surveys with the probability of finding (or missing) an individual sign (in this case tiger scat) (r). In other words, p represents the number of surveys out of the total surveys that are likely to detect presence of tigers in an occupied site, while r represents the proportion of tiger signs that are detected (or missed) in a single survey. The two are not the same i.e. p ≠ r. For example, a survey that detected nine out of 10 signs present or another that detected one sign out of 10 signs are both considered as having 100% detection of tiger presence (p=1) for an occupancy survey, but r for each of these surveys is 0.9 and 0.1 respectively. Thus, detection probability (p) of occupancy surveys is not informative on per capita detection rates (r) of tiger sign. For estimating r the correct approach would be to use a double blind observer experimental design (Buckland et al. 2010, Nichols et al. 2000), where two observers would walk the same trail some distance apart and record observed tiger scat without communicating with each other. The scats being missed by each of them could then be used to estimate the probability of missing scats entirely.

Also, in occupancy surveys all kinds of signs are often used to detect tigers (pugmarks, scat, scrape, rake marks, direct sightings, vocalization, tiger kills, etc). Karanth et al. (2011a) have used both tiger scat and tiger pugmark to detect tigers in a grid for estimating occupancy. Thus, detection probability of occupancy in these surveys is the compounded probability of
occurrence and detection of both scat and pugmark on a single survey which cannot be teased
apart and used as a surrogate for detecting individual scats. From the above it is clear that the use
of occupancy detection probability to parameterize detection probability of tiger scat in the
theoretical model of Gopalaswamy et al. (2015) is wrong. Typically in a double sample survey
the index is measured without an estimate of its detection, by calibrating this potentially biased
index against abundance, double sampling elegantly addresses the issue of detection and other
sources of variability in the index (Conroy and Carroll 2009).

b) Selectively picked references.
Not only do Gopalswamy et al (2015) use an incorrect detection probability (derived for
occupancy studies) in place of a double observer based detection probability for sign intensity for
their theoretical model, they were selective in picking low estimates of detection probability with
high coefficient of variation (CV) from those available in published literature. The estimates of
detection probability p at 1 km segments (0.17) and its CV (1) from Karanth et al (2011a) were
used, claiming that these were the only parameter estimates available. The use of low p and
extraordinarily high CV to suggest that detection of tiger presence for occupancy survey is in
general low and highly variable. These parameters play an important role in subsequent
derivations in the paper. Gopalswamy et al (2015) have ignored other published estimates of
these parameters obtained by sampling large areas and derived by following the same field and
analytical protocols. These publications report far higher p with much smaller CV (Harihara and
Pandav (2012), p = 0.951 SE 0.05; Barbara-Meyer (2013), p=0.65 SE 0.08). The low p and high
CV reported by Karanth et al (2011a) is likely due to poor design and not a norm in detecting
tiger presence. In our experience tigers uses scat, scrape and rakes to advertise their presence and
it is highly unlikely that tiger signs will have such a low detectability unless the population is
very low, survey design is poor, or data is collected by inexperienced/untrained persons.

c) Inappropriate and incorrect data
Throughout the paper the authors have used data and parameters related to tigers published by K.
Ullas Karanth (a co-author on the paper) and colleagues, which they refer to as Index-calibration
experiment – (IC-Karanth). The authors have presented eight paired data points on tiger density
and tiger signs (in fact only scats) in figure 5 of the paper. This graph shows no relationship
between tiger scat encounter rate and tiger density, considered as an empirical test in support of
their theoretical model based only on eight data points. On perusal of the references cited in Gopalaswamy et al. (2015a), we noticed several irregularities which invalidate the use of these data as a scientific experiment to test this relationship. It is relevant to point out that for calibration of any index with abundance as done in a double sampling experimental approach (Eberhardt & Simmons 1987), both index and abundance, should be sampled contemporaneously and over the same spatial extent (paired in time and space). In three data points out of eight presented in figure 5 of Gopalaswamy et al. (2015), tiger signs and tiger density were not collected contemporaneously. Tiger density can fluctuate substantially between years (Karanth et al. 2006) and tiger signs have short persistence time. Yet, the data Gopalaswamy et al. (2015) use for their paired experiment has lags of several years (two to seven years) between estimating tiger density and tiger sign (Figure 1). In particular, the data point from Bandipur has a lag of seven years (density estimated in 1999, scat sampling in 2006), data point representing Melghat has a lag of three years (density estimated in 2002, scat sampling done in 2005) and data point from Pench Maharashtra has a lag of two years (density estimated in 2002, scat sampling done in 2004) (Karanth & Nichols 2000, 2002; Karanth et al. 2004; Karanth & Kumar 2005; Andheria 2006, see supplementary material for relevant sections of these publications). The authors do have concurrent density estimates from one of these sites (Bandipur) with smaller variance (Gopalswamy et al 2012), but curiously have not chosen to use or refer to this. At one data point (Tadoba), an extreme outlier at right corner of figure 5 of Gopalaswamy et al (2015a) (Figure 1), the data on scat encounters does not match the original source (scat encounter rate 3.6/10 km as given in figure 5 of Goplaswamy et al (2015) vs. 1.99/10 km as given in the original source (Karanth & Kumar 2005; but addressed this by mentioning that the original reference was incorrect in a corrigendum to the original paper Gopalaswamy et al (2015b)). Yet two data points (Melghat and Pench Maharashtra) continue to differ in their Fig 5 (Gopalaswamy et al (2015a) from the cited references in the corrigendum Gopalaswamy et al (2015b).

Methods for recording scat encounter rates differed between source reference sites used for IC-Karanth. Andheria (2006) removed all scats encountered on the first sample and discarded them from data analysis, a practice which is not uniformly followed for recording tiger scat encounter rates in other studies. For studies referenced for IC-Karanth, camera trap sampling was done in small areas within larger protected areas for estimating tiger density, whereas tiger scats were collected for studying tiger diet (Karanth & Nichols 2000,
opportunistically from the entire reserve. Any intent of calibrating these tiger scat data to
tiger density obtained through camera trap sampling is not mentioned in any of the original
sources. In the original studies cited by Gopalaswamy et al. (2015) referred to as IC-
Karanth experiment, there seems to be no intent of designing an experiment to evaluate the
relationship between tiger sign encounter rate and tiger density, the sources are unclear if
the scat sampling was done within the same spatial extent as the camera trap survey for
estimating tiger density. The basic premise of a double sampling experimental approach,
wherein data from both samples (index and density) need to be paired in time and space is
violated in the field experiment (IC-Karanth) of Goplaswamy et al (2015) invalidating their
conclusions.

[Figure 1. Here]

d) Variability in tiger capture probability and density estimates from camera trap capture-
mark-recapture.

As with occupancy detection probability, Gopalaswamy et al. (2015) restrict themselves entirely
to 11 estimates of tiger density published by Karanth et al (2004) for their models. On multiple
occasions they point out the highly variable capture probability p and variance associated with
tiger density estimates. In fact, in light of the large number of published tiger density estimates
with higher precision (e.g. 21 estimates in Jhala et al 2011a), these authors should have
considered Karanth et al (2004) estimates as particularly lacking in precision. When, estimates
with large sampling errors are used to guide development of theoretical models it would be
difficult to deduce any relationship between tiger signs and tiger density. Poor precision of tiger
density estimates in Karanth et al (2004) were likely due to poor sampling design and not
something that is inherent in tiger population estimation, e.g. for data presented in Karanth et al
(2004) CV of tiger density increases with increase in sampled area and p decreases with the
sampled area (r =0.4, and -0.63 respectively). Overstating the case of sampling uncertainty can
only do harm to the development and adoption of sound and practical methods.
e) **Repeating non peer-reviewed literature** to advance unsubstantiated claims

Gopalaswamy et al. (2015) claim that the methods followed by Jhala et al. (2011a) have resulted in “improbable estimates of 49% increase in tiger density over 4 years”. Gopalaswamy et al (2015a) do not explain how they arrived at the figure of 49% increase, they cite a letter to Science, commenting on a news article (Karanth et al 2011b), but they have not explained the 49% increase in tiger abundance in this letter as well (Jhala et al. 2011c). The fact is that in 2006 India’s mean tiger population was estimated at about 1400 while in 2010 the estimate was about 1700 but included estimates from some new areas like Sundarbans that were not assessed in 2006. Comparing tiger numbers between common areas sampled in 2006 and 2010 an increase of 17.6% was estimated in four years, or about 4% per year; which is very probable for large carnivores. It is inexplicable to us how Gopalaswamy et al (2015a) arrived at a 49% increase in abundance or why they continue to perpetuate this obviously erroneous inference.

f) **Propaganda that is not consistent with facts**

The paper of Gopalaswamy et al (2015a) is, as the title suggests, about "index calibration experiment" especially referring to estimation of tiger abundance. To this extent the reference to Jhala et al (2011a) that demonstrates a strong relation between tiger sign index and tiger abundance as IC-Jhala and several publications of U. Karanth as IC-Karanth is relevant. Gopalaswamy et al. (2015a) seem to have gone through the methods employed for estimating the status of tigers in India thoroughly (Jhala et al. 2008, 2011b, and 2015), since they have meticulously computed parameters from these reports for their paper. K. U. Karanth is also an author on several chapters in Jhala et al (2015). They should know that national tiger status assessments (Jhala et al 2008, 2010, 2015) were never based on tiger sign index alone. Tiger sign index was one amongst the many ecologically important covariates that included human footprint, prey abundance and landscape characteristics, that were used for modeling tiger density. Yet, the blog of the journal Methods in Ecology and Evolution titled "flawed method puts tiger rise in doubt " states "amongst recent studies thought to be based on this method is India’s national tiger survey " (Grives 2015) which the blog then discredits as being inaccurate based on conclusions of Gopalaswamy et al. (2015a). The fact is India's national tiger survey of
2014 (Jhala et al. 2015) used spatially explicit capture-recapture (SECR) in a joint likelihood based framework (Efford 2012) with covariates of prey abundance, tiger sign intensity, habitat characteristics, and human footprint. The SECR and Joint likelihood analysis are a recent development (Brochers & Efford 2008, Efford 2011) and therefore could not have been used for earlier national tiger assessments which used general linear models (Jhala et al 2008, 2011b).

The misleading reports that subsequently followed in the media had forgotten that the MEE paper by Gopalaswamy et al. (2015a) is a debate on index calibration using double sampling approach (Eberhardt & Simmons 1987) with simple linear regression and not about national tiger status assessment. The 2014 national tiger status assessment was based on photo-captures of 1506 individual tigers, capture-histories of these were subsequently modeled in SECR with covariates of prey, habitat and human impacts to estimate 2226 (SE range 1945 to 2491, >1.5 year old) tigers from across India (Jhala et al 2015). This amounts to 68% of the total tiger population being photo-captured and 77% (1722; 95% CI 1573 to 2221 tigers) of the total tiger population being estimated by capture-mark-recapture without any extrapolation using covariates/indices. By muddling index calibration with the national tiger survey in the paper (Gopalaswamy et al. 2015) and in all subsequent press releases and interviews Dr. Ullas Karanth and coauthors incorrectly use the Gopalaswamy et al. (2015a) paper results (which are themselves highly questionable) to discredit the national tiger survey results as being inaccurate (Bagla 2016, Chauhan 2015, Croke 2015, Grives 2015, Karanth 2015, 2016, Rohit 2015, Sinha and Bhattacharyal 2015, Vaughn 2015, and Vishnoi 2015) and mislead the readers.

Peer reviewed publications form the basis for advancement of science and are often cited and used as a basis from which to move ahead. Indeed, the Gopalaswamy et al. (2015a,b) paper has been subsequently cited in papers addressing methodological reviews, advances and policies (Darimont et al 2018, Hayward et al. 2015), abundance estimation papers (Broekhuis and Gopalaswamy 2016, Caley 2015, Elliot and Gopalaswamy 2017, Falcy et al 2016, Mahard et al 2016) and in some Masters and Ph.D. thesis (Walker 2016, Moorcroft 2017). Published scientific literature can have errors, these can occur through negligence of scientists or deliberate misleading of science (Macilwain 2014), and can pass the peer review process due to ignorance, poor diligence, or vested interest (Parsons & Wright 2015). Mistakes in published science should be corrected expediently, as these are detrimental to the scientific progress in the specific field and propagate a wrong basis for further research. In our opinion, Gopalaswamy et al.
results are misleading, due to inappropriate scientific process and data, and have therefore not contributed to the wider debate on the usefulness of double sampling (Eberhardt & Simmons 1987; Pollock et al. 2002) for large-scale animal surveys. We stress that landscape scale surveys need to be a blend of robust statistical design and analysis that are pragmatic (economic and logistically possible) to achieve. The national tiger surveys of India (Jhala et al. 2008, 2011, 2015) have striven to keep pace with modern advancement in animal abundance techniques and analysis and have used robust statistical tools available within the constraints of large-scale data coverage, resources, and timeframe. The concept and philosophy of double sampling (Cochran 1977) forms the basis for modern statistical and analytical approaches that infer relationships between actual abundance and counts, indices, and covariates. The family of general linear models, generalized additive models (Zuur et al. 2009), joint likelihood (Conroy et al. 2008), SECR with habitat covariates (Efford and Fewster 2013), and SECR joint likelihood (Chandler and Clark 2014) take the relationship between an index/covariates and absolute abundance to various levels of analytical complexity. There seems to be some agreement on the best analytical approach to use for landscape scale abundance estimation of tigers between Gopalaswamy et al. (2015a) and us (Jhala et al. 2015). Gopalaswamy et al. (2015a) recommend using the joint likelihood approach, while the tiger status assessment for India for the year 2014 used spatially explicit joint likelihood with camera-trap data of tigers, and covariates of tiger sign index, prey abundance and human footprint indices (Jhala et al. 2015). Yet, we stress the relevance and importance of first exploring relationships of abundance with indices and covariates, based on sound ecological logic before attempting complex statistical analysis, and refrain from putting the proverbial cart (statistical) before the horse (ecology) (Krebs 1989).

Acknowledgements:

We acknowledge Rashid Raza for painstakingly retrieving old reports to verify data used for IC-Karanth and helping draft this paper. We thank S. Dutta, S. Bist, Bipin C., and V. Kolipakam for their comments on the manuscript.
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Figure 1. Recreation of figure 5 from Gopalaswamy et al. (2015) highlighting the data discrepancies in the index-calibration experiment. The names of tiger reserves from central Indian landscape and Western Ghat landscape, where sampling was done are mentioned. MR refers to the State of Maharashtra, and MP refers to the State of Madhya Pradesh.
Recreation of figure 5 from Gopalaswamy et al. (2015) highlighting the data discrepancies in the index-calibration experiment. The names of tiger reserves from central Indian landscape and Western Ghat landscape, where sampling was done are mentioned. MR refers to the State of Maharashtra, and MP refers to the State of Madhya Pradesh.
Tiger Density (tigers per 100 sq km) vs. Tiger Sign Encounters (Scat per 10 km)

- **Kaziranga**: 16.00 (Tiger Density), 0.00 (Tiger Sign Encounters)
- **Nagarahole**: 12.00 (Tiger Density), 2.00 (Tiger Sign Encounters)
- **Bandipur**: 14.00 (Tiger Density), 2.00 (Tiger Sign Encounters)
- **Kanha**: 14.00 (Tiger Density), 2.00 (Tiger Sign Encounters)
- **Melghat**: 6.00 (Tiger Density), 3.00 (Tiger Sign Encounters)
- **Pench-MP**: 3.00 (Tiger Density), 2.00 (Tiger Sign Encounters)
- **Pench-MR**: 8.00 (Tiger Density), 3.00 (Tiger Sign Encounters)
- **Tadoba**: 4.00 (Tiger Density), 1.00 (Tiger Sign Encounters)

Key Notes:
- **Kanha**: Mismatch between Fig & source ref, 7 year gap between density & scat.
- **Bandipur**: Mismatch between Fig & source ref, 2 year gap between density & scat.
- **Pench-MR**: Mismatch between Fig & source ref, 2 year gap between density & scat.
- **Melghat**: Mismatch between Fig & source ref, 3 year gap between density & scat.

Clarity and Coherence:
- The diagram clearly illustrates the comparison of tiger density and sign encounters across different areas.
- Labels and annotations provide context for each location and key differences observed.

Technical Accuracy:
- The axis labels and data points are clearly marked.
- The data trend lines are visually represented to show the relationship between tiger density and sign encounters.

Potential Questions:
- How do the observed gaps in data between tiger density and sign encounters affect conservation efforts?
- What are the implications of mismatches noted in the data for tiger population monitoring and management?