Local shifts in floral biotic interactions in habitat edges and their effect on quantity and quality of plant offspring

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Abstract. Spatial shifts in insect fauna due to ecological heterogeneity can severely constrain plant reproduction. Nonetheless, data showing effects of insect visit patterns and intensity of mutualistic and/or antagonistic plant–insect interactions on plant reproduction over structured ecological gradients remain scarce. We investigated how changes in flower-visitor abundance, identity and behaviour over a forest-open habitat gradient affect plant biotic interactions, and quantitative and qualitative fitness in the edge-specialist Dianthus balbisii. Composition and behaviour of the insects visiting flowers of D. balbisii strongly varied over the study gradient, influencing strength and patterns of plant biotic interactions (i.e. herbivory and pollination likelihood). Seed set comparison in free- and manually pollinated flowers suggested spatial variations in the extent of quantitative pollen limitation, which appeared more pronounced at the gradient extremes. Such variations were congruent to patterns of flower visit and plant biotic interactions. The analyses on seed and seedling viability evidenced that spatial variation in amount and type of pollinators, and frequency of herbivory affected qualitative fitness of D. balbisii by influencing selfing and outcrossing rates. Our work emphasizes the role of plant biotic interactions as a fine-scale mediator of plant fitness in ecotones, highlighting that optimal plant reproduction can take place into a restricted interval of the ecological gradients occurring at forest edges. Reducing the habitat complexity typical of such transition contexts can threaten edge-adapted plants.

Keywords: Dianthus; dichogamy; edge effect; inbreeding depression; plant mating systems; pollen limitation; self-fertilization.

Introduction

Plant biotic interactions (i.e. plant functional relationships with mutualistic and antagonist organisms; PBIs hereafter) are a key driver of population processes in angiosperms, the largest group within the plant kingdom. In flowering plants, PBIs are mainly founded on associations with insects, which may act as pollinators or preda-
As far as mutualistic relationships are concerned, 308 006 flowering plant species, accounting for 87.5% of the whole angiosperm diversity, show animal-mediated pollen transfer (Ollerton et al. 2011). The relevance of plant–pollinator interactions in preserving viable plant populations is emphasized by historical connections between insect and plant extinction patterns (Pauw and Hawkins 2011). Indeed, inadequate pollen delivery (i.e., pollen limitation) is recognized to limit plant fitness (Aguilar et al. 2006; Ghazoul 2005; Knight et al. 2005) and, then, to affect angiosperm evolution (Harder and Aizen 2010). Pollen limitation (PL) can derive either from quantitative or qualitative inadequateness of the pollen reaching the stigma (Wilcock and Neiland 2002). In the first case, PL originates from insufficient pollen transfer, which is typically due to the scarcity of pollinators (Ashman et al. 2004). Instead, qualitative PL is due to a poor quality of the pollen admixture deposited onto stigmas (Amat et al. 2011; Arceo-Gomez and Ashman 2014). Variations in abundance, composition and behaviour of pollinating fauna may represent a primary source of both quantitative and qualitative PL (Gómez et al. 2010). Such insect variations can rise from habitat changes (Tottland 2001; Chalcoff et al. 2012), human disturbance (Brys and Jacquemyn 2012) and local structuring of plant populations (Zorn-Arnold and Howe 2007; Le Cadre et al. 2008).

Although much attention has traditionally paid to plant–pollinator relationships, recent works highlights the need of integrating studies on both plant–pollinator and plant–herbivore interactions in investigations on PBIs (Lucas-Barbosa 2016). Herbivory often exerts major constraints on plant reproduction (Cariveau et al. 2004; Ashman and Penet 2007). For instance, florivory may alter the efficiency of relationships between plants and pollinators. This may increase PL in plant populations, due to the reduced pollen transfer caused by pollen consumption (Hargreaves et al. 2009; Harder and Aizen 2010), and to herbivory-induced variations in plant traits involved in pollinator attraction, like floral display (Penet et al. 2009; Söber et al. 2010) and scent (Lucas-Barbosa et al. 2011, 2016). Such interplay between mutualistic and antagonistic relationships has also important evolutionary implications, often resulting in selection conflicts on traits mediating insect attraction or avoidance (Gómez 2003, 2005).

Because habitat variations influence patterns of insect presence, abundance and behaviour (Tottland 2001; Caruso et al. 2003), plant processes relying on PBIs can show relevant spatial variations (Gómez et al. 2010). These variations may be particularly evident in proximity of habitat edges. Indeed, the switch between different habitats generates transition zones with emergent ecological properties recognized as ecotones (Risser 1995). In landscapes subjected to a long-lasting human impact, most of spatial ecological heterogeneity (and ecotones) derives from the alteration of pristine habitats like forests. This contributes to form a kaleidoscope of transition habitats, which change in space and time depending on local patterns of disturbance and vegetation dynamics (Dovciak and Brown 2014). Studying ecological responses at forest edges (i.e., edge effect) is a key topic in landscape ecology and biodiversity conservation (Forman 1995; López-Barrera et al. 2007; Hadley and Betts 2012). Indeed, understanding how ecological processes vary near forest edges is crucial to evaluate how landscape-level dynamics, such as fragmentation, affect biodiversity (Ries et al. 2004). Overall, species interactions are retained one of the main mechanisms influencing biodiversity patterns in habitat edges (Ries et al. 2004). Accordingly, plant population processes depending on PBIs (i.e., herbivory, pollination and dispersal) are influenced by the proximity to forest edges (Chacón and Armesto 2006; López-Barrera et al. 2007). The extent of such kind of edge effect depends on local levels of ecological heterogeneity (Sarlov-Herlin 2001; López-Barrera et al. 2007), and disturbance (Bhattocharya et al. 2003; Huang et al. 2009). Because the functioning of PBIs is influenced by the size and spatial configuration of habitat patches suitable for plant species (Lennartsson 2002; Verbouven et al. 2014), the influence of habitat heterogeneity on PBIs becomes more relevant in highly fragmented landscapes (Hadley and Betts 2012).

Studying PBIs patterns in the proximity of forest edges would improve plant conservation in complex ecological mosaics (Hadley and Betts 2012), as those found in the Mediterranean region. In this area, due to the fragmentation of natural habitats, habitat edges occupy a conspicuous portion of current landscapes, and are expected to augment in the future, especially in lowlands (Falcucci et al. 2007). Consequently, such contexts will represent the theatre for evolutionary and conservation challenges of an increasing number of plants. Nonetheless, the management requirements of such anthropogenic ecotones are often neglected. To date, there is a substantial scarcity of field research investigating how small-scale habitat variations affect PBIs (including plant–pollinator and plant–herbivore interactions), and then plant reproductive performance along structured ecotone gradients.

We evaluated the influence of ecological heterogeneity on PBI-driven population processes in Dianthus balbisii Ser., an edge-specialist inhabiting ecotones between woody and open habitats in Mediterranean low lands. Field work aimed to (i) identify potential pollinators and herbivores of D. balbisii; (ii) evaluate the aptitude of different pollinators versus intra- or inter-plant visits; (iii) check for visit frequency variations along a gradient of light intensity (assumed as a proxy of habitat...
structure); (iv) perform floral manipulations to compare the outcome of natural pollinations with that of forced crossing and forced selfing along the studied ecological gradient. Such data were used to ask the following questions: (i) How did insect patterns change along the studied gradient? (ii) Did variations of insect visit patterns cause functional shifts in PBIs (i.e. pollination likelihood, herbivory rate, amount and quality of pollen transfer)? (iii) Did such shift in PBIs influence quantitative and qualitative components of plant reproductive fitness along the studied ecological gradient?

By answering such questions, we would provide basic insights on the relationships between PBIs and plant reproductive fitness in anthropogenic ecotones, an integral component of complex landscapes.

Methods

Study species and study area

The perennial *Dianthus balbisii* Ser. has a woody stock bearing up to 20 flowering stems ending with a multi-flowered head (2–20 flowers × head); the corolla is reddish, by distinguishing *D. balbisii* from the closely related *D. guiae* Janka (Peruzzi and Gargano 2006). The flowering season is from late May to late June. The distribution range of *D. balbisii* embraces the Central Mediterranean area, where this species occurs in hilly areas at the border between woody and open xeric communities (Pignatti 1982). As in several congeneric species (Erhardt 1988), the plant is self-compatible, and shows a protandrous hermaphroditism with the possible occurrence of female flowers (i.e. gynomonoicism and gynodiocism). *Dianthus* species have rather specialized pollination syndromes, involving Lepidoptera as main pollinators (Erhardt 1988). Floral traits and architecture (i.e. reddish flowers with short calix grouped in dense heads) suggest that *D. balbisii* belongs to the carnations pollinated by diurnal butterflies (Erhardt 1988; Block et al. 2006). Field observations (i.e. daily flower monitoring) also support that in this species most of pollinations occur during the day (D. Gargano pers. obs.). In the study area (Calabria, Southern Italy, *N* 39°33′; *E* 16°22′), *D. balbisii* occurs near the border of forests dominated by deciduous oaks, and in adjacent garrigue-like communities.

Sampling individuals and structuring the habitat gradient

The overall study population accounted approximatively for 150 plants; these were rather uniformly distributed over a forest ecotone (~150 × 50 m) encompassing two adjacent sites in a highly anthropized area. In spring 2014, 83 individuals of *D. balbisii* were permanently marked before flowering by selecting plants that were at least 5 m apart. From the whole sample, 40 plants were used to collect data about pollination and herbivory. The remaining 43 plants were employed in experimental pollinations. In this way, possible interference of floral manipulation procedures (e.g. bagging) with PBIs were avoided.

The ecotone structure was determined based on measures of canopy closure. This parameter was chosen because variations in habitat canopy are recognized to influence both abiotic (i.e. temperature and soil moisture) (Joffre and Armesto 2006; Lopes-Barrera et al. 2007) and biotic (Totland 2001; Chacon and Armesto 2006) drivers of plant fitness. The position over the ecotone of each sampled plant was determined by using light intensity (*LI*) as a proxy of habitat canopy closure. A direct measure of incident light was chosen because it can be effective in recording small-scale canopy variations near the ground surface, as congruent with the size of the study species. Direct solar radiation measurements can also help in interpreting patterns of pollinator activity (Kasting and Weber 2001). Moreover, this method allows many replications to obtain more stable average values. *LI* was measured on the surface of the flower head by a luxometer (Panlux electronic 2, Gossen, Nürnberg). For plants that had up to three flowering stems the light intensity was measured on all stems, and the recorded values were averaged. Instead, for plants with many flowering stems, three measures were taken on the three tallest stems, and the recorded values were then averaged. All measures were carried out after the canopy of deciduous trees completed its development. To reduce the effect due to the moment in which they were taken, all the measurements were repeated during two consecutive sunny days (12–13 June) from 12.00 a.m. to 2.00 p.m.

*LI* measurements were used to structure a gradient of light intensity encompassing five classes hosting a comparable amount of sampled individuals: (i) *LI* < 500 Klux (No. plants = 16), (ii) 500 < *LI* < 1500 Klux (No. plants = 17), (iii) 1500 < *LI* < 2500 Klux (No. plants = 17), (iv) 2500 < *LI* < 4000 Klux (No. plants = 17), (v) *LI* > 4000 Klux (No. plants = 16). Because canopy variations regulate spatial resource patterns in temperate habitats (Joffre and Armesto 1993), plants occurring in the same *LI* interval were assumed to have a comparable amount of available resources as they occurred under similar canopy conditions.

Visiting insects

Individual patch size (no. of stems, no. of buds per stem and per plant) was estimated for the 40 plants (8 × *LI* class) selected to monitor insect visits.
Observations on insects visiting *D. balbisii* were carried out over four consecutive days (from 12 June to 16 June) in the middle of the species’ blooming season. Overall, they accounted for 560 min of direct field observations for recording each insect visit on each marked plant. Such observations were organized in trials of 20 min, which were carried out during fixed intervals in the morning (from 10.30 a.m. to 12.30 a.m.) and in the afternoon (from 3.00 p.m. to 5.00 p.m.). Plants were subjected to 28 observation runs (16 in the morning and 12 in the afternoon). To minimize disturbance to insects, the observers remained seated some meter apart the plants. Due to the gradual blooming typical of the study species, each observation trial was preceded by the annotation of the amount of flowers opened on the plants. This allowed us to relate visit patterns to the effective flower display (EFD = no. of opened flowers on the plant at the visit time).

When possible, the identity of insect taxa was assessed in the field. Unknown insects visiting *D. balbisii* flowers were identified in laboratory with the help of expert entomologists based on images and specimens collected in the field. Taxa unidentified at the specific level were assigned to higher taxonomic ranks (i.e. genus and family). The observed insects were categorized in four functional categories. Firstly, they were distinguished into pollinators or herbivores based on literature knowledge and field observations. Butterflies were always considered as pollinators as they are the main pollinators of *Dianthus* species (Erhardt 1988). Also *Bombylius major* L. (Diptera), that frequently visited flowers of *D. balbisii*, was included among pollinators. Indeed, bee-flies are recognized as pollinators in many plant species with tubular flowers (Kastinger and Weber 2001). Then, the primary pollinators (the lepidopteran *Thymelicus sylvestris* Poda and *B. major*) were separated from secondary pollinators (other lepidopterans, OTH) based on visit frequency. Furthermore, *T. sylvestris* and *B. major* were assigned to two distinct categories (named, respectively, THY and BOM) because of relevant functional differences (i.e. diverse preferences with respect to the investigated gradient, diverse aptitude to within- and among-plant visit). Finally, weevils and ants were classified as florivores (McCall and Irwin 2006; Penet et al. 2009) or pollen and nectar robbers (Soper Gorden and Adler 2016), respectively. Since weevils were widely dominant, all florivores were included into a single category (HER). HER also included pre-dispersal predators that were identified in laboratory based on the caterpillars found in fruits of *D. balbisii* collected in the field. Joining florivores and pre-dispersal predators into a single unit was justified by the very low amount of seeds produced by flowers that experienced florivory (see below).

To estimate the potential for cross- and self-fertilization of each visiting taxon based on its behavioural aptitude, visits were classified as (C) (visits among individuals, potentially allowing for cross-pollination), or (G) (visits within individual, potentially allowing for geitonogamy). The (C) tag was used when insects visited an individual of *D. balbisii* and then moved away. Differently, the (G) tag applied to visits of insects coming from other flowers of the same plant (in other words, multiple consecutive visits on a same plant qualified as G).

**Flowering regime**

The 40 plants were monitored daily over 3 weeks (from 14 June to 5 July), by annotating amount and sex of open flowers. Since the species is protandrous, the hermaphrodite flowers were distinguished into males (i.e. flowers bearing only emergent stamens), and hermaphrodites (i.e. flowers bearing both stamens and stigma branches). Finally, male-sterile flowers (i.e. bearing only stigma branches) were classified as female. Flower sex data were used: (i) to quantify the frequency of male sterility within the population, and (ii) to estimate pollination frequency over the study gradient (i.e. in protandrous species, large accumulations of flowers in the hermaphrodite stage can indicate slow pollination processes).

During such field trials we also recorded flowers that resulted pollinated (i.e. flowers with closed corolla) or showed signs of florivory (i.e. petals, stamens and/or stigma branches eaten by herbivores). This work consisted in 2737 field observations involving 656 flowers.

**Floral manipulations**

Fertility of perennial plants can be affected by several maternal traits (e.g. age, size, history, genetics and environmental conditions). This can bias fitness comparisons among pollination treatments carried out on different individuals. To reduce such bias, floral manipulations were designed to obtain matched offspring for various pollination protocols (see below) from each experimental plant. To this end, from the initial 43 individuals, plants bearing <5 flowers were not manipulated, and the final amount of treated plants became 30 (6 × *LI* class).

The pollination experiments performed on these plants involved at least one flower for each of four treatments: forced crossing (No. = 80), forced selfing (No. = 76), autonomous selfing (No. = 74) and free pollination (No. = 80). Except for the sample used for free pollination, the flowers were isolated at the early bud stage by bagging 1–2 inflorescences per plant. In flowers used for forced cross-pollination, anthers were gently
removed with a dissection forceps before the emergence of stigma to avoid occasional selfing. Hand-pollinations (i.e. forced crossing and forced selfing) were carried out by brushing one dehiscent anther per stigma branch; this would allow for fertilizing all ovules (<100) usually found in the ovaries of D. balbisii. The stigma was retained receptive when it was well developed and appeared glutinous and papillose. Self-pollen was taken from other flowers of the same plant. Pollen used in forced-crossing was taken from >30 plants, each distant >5 m from the recipient plant. Occurrence and extent of autonomous self-pollination was evaluated by leaving untreated 1–2 bagged flowers per plant. As the last flower in the head closed petals or withered, the bag was removed to allow fruit development under natural conditions.

**Fitness components**

All flowers left free were used to investigate rates of herbivory, pollination and seed set under natural pollination regime. At the end of the blooming season all fruits occurring on the marked plants were collected. In the laboratory, fruits were checked for occurrence of pre-dispersal predation (e.g. presence of caterpillars, remnants of seeds eaten by herbivores). In these way fruits were distinguished into three categories: predated, aborted and fertilized. This screening accounted for 1185 fruits from 40 plants. Except for three cases (on 571), no seeds were produced by flowers that experienced florivory or pre-dispersal predation. Therefore, we annotated for each flower the occurrence of herbivory or pollination, and in the latter case the number of seeds.

The seeds contained in each fruit were counted and, then, a sample of 450 seeds was germinated to assess seed viability and early seedling survival rate. At the beginning of the autumn seeds were sowed on wet paper in Petri dishes at 18 °C. The resulting seedlings were transferred into 4 × 4 cm pots containing a mixture of peat moss and compost with presence of pumice fragments. Subsequently, 30-day old seedlings were transplanted into 20 × 17 cm pots filled with brown soil and subjected to periodical irrigation and fertilization (20 ml of a mixture of ammonium (6 %), phosphorous (5 %), and potassium (7 %) diluted in H2O was monthly supplied). The plants were maintained in glasshouse for 6 months at Botanic Garden of the Università della Calabria.

**Data analysis**

The efficacy of the sampling effort in capturing information on insects visiting D. balbisii was evaluated by a rarefaction procedure on sample-based abundance data (Colwell et al. 2012). To this end it was applied the rarefaction function provided by EstimateS 9.1.0 (Colwell 2013).

The relationships between visit frequency (no. of visits × plant patch × observation run) and EFD were checked by a Pearson correlation test. Variations of visit frequency across levels of light intensity were checked by a Generalized Linear Model (GLM), by considering the LI class as fixed factor and EFD as covariate. The association between visitor type and light intensity, and between visitor type and visit type was evaluated by cross-tabulation analyses using chi-square as a measure of association.

Variations of individual patch size (i.e. no. of stems per plant, no. of buds per plant and no. of buds per stem) across intervals of light intensity were analysed by one-way Anova. Patterns of sex ratio were evaluated by calculating from daily data the individual hermaphroditic rate (no. of hermaphrodite flowers/total no. of open flowers per plant) and the individual female rate (no. of female flowers/total no. of open flowers per plant). Then, differences among intervals of light intensity in average individual hermaphroditic and female rates were checked by one-way Anova.

The strength of PBI was evaluated based on the variations of herbivory, pollination chance and seed set over the five LI classes. In the analyses, herbivory accounted for both florivory and pre-dispersal predation, because they had an equivalent effect of the reproduction of D. balbisii (i.e. both caused absence of seed-set). To include the effect of plant traits involved in insect attraction on herbivory and pollination likelihood, we comprised in such analyses the inflorescence size (i.e. the amount of flowers occurring in the same head of each studied fruit). Herbivory was a binary variable indicating presence or absence of florivory and pre-dispersal predation. The pollination chance also consisted in binary data indicating occurrence or absence (caused by lack of visits, herbivory or abortion) of fertilization. Variations of herbivory and pollination chance among LI classes were then analysed by GLM, including the LI class as fixed factor and the inflorescence size as covariate. Quantitative fitness analyses were based on seed set (i.e. no. of seeds per fruit) comparisons between naturally pollinated and manipulated plants. Flowers that were not pollinated were considered by assigning a seed set = 0. Seed set in not-predated flowers (570 flowers from 40 plants) was analysed by one-way Anova to estimate seed set variations among the 5 LI classes. To evaluate the effect of qualitative pollen limitation on seed production, a GLM was used to analyse the seed set of 75 fruits per each of three treatments: forced-crossing, forced selfing, natural pollination (3 fruits × 5 plants × 5 LI classes × 3 treatments) using LI class and pollination treatment as fixed factors.
Further qualitative fitness components (i.e. seed germination and survival in 6-month-old seedlings) were estimated by considering the sample of 450 seeds (6 seeds \( \times 5 \) plants \( \times 5 \) LI classes \( \times 3 \) pollination treatments) obtained by plants under free-pollination, forced crossing and forced selfing. Autonomous self-fertilization was not considered because it was ineffective in most of cases. Seed germination and seedling survival were treated as binary variables, and were analysed by GLM (using LI interval and pollination treatment as fixed factors). GLM analyses on binary variables were performed by using the LOGIT link function. Response variables accounting for count data (i.e. visit frequency, no. seeds per fruit) were not normally distributed. Therefore, in GLM these variables were analysed by using a LOG link function, which is appropriate for data represented by non-negative integer values. Because floral manipulations (i.e. pollen supplementation) can promote resource reallocation within plants, this may result in a higher pollen limitation in free-pollinated flowers of manipulated plants (Knight et al. 2006; Wesselingh 2007). The possible effects of such reallocation processes were tested by a \( t \)-test on independent samples on the seed set of 160 free pollinated flowers, 80 were randomly selected from the 30 manipulated plants, and 80 from 30 non-manipulated individuals within the same LI interval.

All these analyses were performed in SPSS 24 for Windows (SPSS, Chicago, IL, USA).

**Results**

**Patterns of visit abundance, visitor type and behaviour over the studied gradient (question 1)**

In average, we recorded 4.7 \( \pm \) 1.9 insect taxa per observation run. The rarefaction function suggested that the observation effort provided a reliable estimation of the insect fauna interacting with the flowers of *D. balbisii* (Fig. 1).

GLM analyses indicated that light intensity influenced visit frequency (Wald \( \chi^2 = 29.46, P < 0.001 \)). More than 71% of documented visits occurred under a light intensity exceeding 2500 Klux (Fig. 2A). The effective floral display (EFD) also significantly influenced the visit frequency (Wald \( \chi^2 = 24.05, P < 0.001 \)). Accordingly, visit frequency was significantly related to EFD (\( r = 0.389; P < 0.001; \text{No.} = 384 \)). However, the significant interaction between LI and EFD (Wald \( \chi^2 = 15.89, P = 0.003 \)) suggested that the effect of EFD on visit frequency varied along the ecological gradient.

The list and frequency of recognized insects, along with their putative functional role, is provided in Table S1 [see Supporting Information]. Among pollinators, *T. sylvestris* (THY) was responsible for 53.2% of visits, followed by *B. major* (BOM = 16.9%), and a mixture of other butterflies (OTH = 5.6%). The remaining visits were due to herbivores (HER = 24.3%). Among the herbivores, florivores (85.5%) were much more frequent than pollen/nectar eaters (14.5%) [see Supporting Information — Table S1]. The frequency of the different visitors varied among intervals of light intensity (Wald \( \chi^2 = 86.389, P < 0.001 \)). The relative visit frequency of THY ranged between 0.28 (LI \( < 1000 \) Klux) and 0.73 (2500 \( < \) LI \( < 3500 \) Klux), and then decreased at higher light intensities (Fig. 2B). Instead, BOM was the most frequent visitor in the lowest LI class (Fig. 2B). The proportion of visits due to HER varied from 0.08 to 0.36, with lower frequencies at intermediate LI intervals (Fig. 2B). The remaining putative pollinators had much lower visit frequency (<0.1 in most LI intervals) (Fig. 2B).

Different visitors were significantly associated to different movement patterns (i.e. among- and within-plant) (Wald \( \chi^2 = 18.34, P < 0.001 \)). Most of within-plant movements were due to BOM (50.7%) and HER (69.7%). Instead, movements among plants prevailed in OTH (60.9%) and, especially, in THY (76.0%).

**Relating insect visit and flowering regime to patterns of PBIs (question 2)**

Patch size traits differed among intervals of light intensity (Table 1). The no. of buds per plant (\( F = 4.948, P < 0.001; \text{N} = 40 \)), the no. of stems per plant (\( F = 3.454, P = 0.02; \text{N} = 40 \)), and the no. of buds per stem (\( F = 4.321, P < 0.001; \text{N} = 40 \)) significantly varied over the gradient, reaching higher values in sunny places. The average individual amount of open flowers found in daily recognitions decreased toward high LI intervals (Fig. 3A), while the rate of florivory showed an opposite pattern (Fig. 3B). Daily average values of individual hermaphroditic rate strongly differed over the LI gradient (\( F = 15.117, P < 0.001; \text{N} = 40 \)), revealing larger scores at lower LI (Fig. 3B). This suggested that, at lower LI, flowers required a longer time to be visited, as congruent with general patterns of insect visit. On the contrary, individual female rate (Fig. 3B) did not reveal significant patterns (\( F = 0.167, P = 0.955; \text{N} = 40 \)).

A variation of PBIs along the studied gradient was confirmed by GLM. Indeed, light intensity, inflorescence size and their interaction) had significant effects on herbivory and pollination chance (Table 2). Pollination increased from low to intermediate intervals of light intensity and then decreased at higher LI, while herbivory showed an opposite pattern (Fig. 4).

**Quantitative and qualitative consequences on plant fitness (question 3)**

The no. of seeds per fruit varied significantly over the gradient (\( F = 14.439, P < 0.001 \)), showing a decline in higher LI intervals (Fig. 5A).
Among manipulated plants, 25 individuals (5 × LI class) produced matched fruits for forced cross-pollination, forced self-pollination and free-pollination. Instead, many flowers and plants failed to produce seeds by autonomous selfing (<14% of flowers produced fruits containing in average 2.45 ± 6.88 seeds). The seed-set of free-pollinated flowers on non-manipulated plants did not differ from free-pollinated flowers on plants subjected to experimental pollinations (t = 1.542, df = 160, P = 0.098). The two way ANOVA on manipulated flowers evidenced substantial differences in seed set among forced selfing, forced crossing and free pollination (Table 3). The lower seed set was found in flowers left to free pollination. In 14% of cases they did not set fruits, and their average seed set (14.75 ± 7.78) was lower than those found in flowers treated with forced selfing (18.05 ± 6.56) and forced crossing (23.41 ± 10.29). LI had significant effect on seed set patterns also in experimentally pollinated flowers (Table 3). The no. of seeds obtained by free pollination was lower than in flowers treated with forced crossing (average difference = −0.866, P < 0.001, N = 150) and forced selfing (average difference = −0.279, P = 0.003, N = 150). As shown in Fig. 5A, seed set from free pollination overcame forced selfing in two LI intervals (LI3: difference = 6.23, P < 0.001, N = 30; LI4: difference = 3.21, P = 0.03, N = 30). In the same LI intervals, the difference between free pollination and hand cross-pollination remained significant but it was less
Table 1. Average values (±SD) of patch size and effective floral display (no. of opened flowers on a plant at the time of insect visit) across classes of light intensity (LI); EFD, effective floral display.

| LI class | Stems × plant | Buds × plant | Buds × stem | EFD  |
|----------|----------------|--------------|-------------|------|
| LI < 500 Klx | 5.30 ± 3.6  | 28.50 ± 26.4  | 4.90 ± 2.2  | 7.57 ± 3.8  |
| 500 < LI < 1500 Klx | 5.75 ± 3.3  | 37.50 ± 31.8  | 5.75 ± 2.5  | 9.08 ± 5.1  |
| 1500 < LI < 2500 Klx | 5.67 ± 5.7  | 43.67 ± 33.3  | 11.94 ± 8.7 | 20.29 ± 13.4 |
| 2500 < LI < 4000 Klx | 6.50 ± 7.3  | 50.50 ± 48.9  | 9.05 ± 6.3  | 12.13 ± 8.2  |
| LI > 4000 Klx | 11.60 ± 11.3 | 61.80 ± 57.3  | 5.80 ± 0.8  | 7.44 ± 3.4  |

Figure 3. Status and sex ratio of the flowers daily recorded over the gradient of light intensity. (A) Relative proportion of open flowers (white bar) and of flowers pollinated or damaged by herbivory (grey bar). (B) Proportion of flowers found in hermaphroditic (stripped bar) or male (dotted bar) stage, and fraction of female flowers (black bar).

Table 2. Effect of light intensity (LI) and inflorescence size (IS) on likelihood of herbivory and pollination.

| Source of variation | Dependent variable | Wald chi² | P      |
|--------------------|-------------------|-----------|--------|
| LI                 | Herbivory         | 40.956    | <0.001 |
|                    | Pollination       | 37.540    | <0.001 |
| IS                 | Herbivory         | 35.785    | <0.001 |
|                    | Pollination       | 26.916    | <0.001 |
| LI × IS            | Herbivory         | 13.977    | 0.007  |
|                    | Pollination       | 10.000    | 0.04   |

Figure 4. Relative proportion of pollination and herbivory over the gradient of light intensity. Light grey bar, proportion of pollinated flowers. Dark grey bar, proportion of flowers subjected to herbivory.
pronounced (LI3: difference = 3.93, P = 0.046, N = 30; LI4: difference = 3.70, P = 0.015, N = 30).

LI and pollination treatment influenced both germination and survival (Table 4). Differences in survival were higher than variations in germination (Fig. 5B, C). Overall, germination rate in cross-pollinated flowers was significantly higher to both free pollinated (83.3 % vs. 72.0 %, \( P = 0.024 \)) and self-pollinated ones (83.3 % vs. 66.0 %, \( P = 0.001 \)). On the contrary, the difference between free pollination and forced selfing was not statistically significant (\( P = 0.231 \)). Seedling survival revealed a similar pattern, although the amount of survived plants was generally lower compared to germination rates (cross-pollination = 75.3 %; free-pollination = 64.7 %; self-pollination = 52.0 %). The differences between cross-pollination and forced selfing (\( P < 0.001 \)) were higher than between cross-pollination and free pollination (\( P = 0.048 \)). Instead, differences between free pollination and forced self-pollination were marginally significant (\( P = 0.051 \)). At intermediate LI intervals (LI = 3–4), fitness differences between cross- and free-pollinations were not significant (\( P = 0.456 \) and \( P = 0.527 \) for germination and survival, respectively) (Fig. 5B, C). Instead, at highest LI the fitness related to free pollination approached that of self-pollination for both germination (\( P = 0.229 \)) and survival (\( P = 0.171 \)) (Fig. 5B, C).

Table 3. Effect of light intensity and pollination treatment on the seed set of manipulated flowers.

| Source of variation                           | F        | P        |
|----------------------------------------------|----------|----------|
| Light intensity                              | 26.489   | <0.001   |
| Pollination treatment                        | 55.292   | <0.001   |
| Light intensity × Pollination treatment      | 8.164    | <0.001   |

Figure 5. Fitness traits patterns across intervals of light intensity and pollination treatments: no. of seeds per fruit (A), seed viability (B) and seedling survival (C). Solid circles, hand-crosspollination; solid squares, forced-selfpollination; empty circles, free-pollination.

Discussion

Linking ecological heterogeneity to patterns of visit frequency and type (question 1)

Spatial ecological heterogeneity can promote substantial small-scale effects on insect webs (Totland 2001; Caruso et al. 2003). Accordingly, our data evidenced substantial variation in the overall and relative frequency of insect visits over the investigated ecological gradient. The individuals of D. balbisii occurring in sunny places experienced high visitation rate, confirming the general expectation that, in forest-edge contexts, insects prefer the more productive open patches (Chacón and Armesto 2006). However, we also recorded relevant variation in the relative visit rate of different insects over the investigated gradient. Substantially, this resulted in three major patterns: (i) infrequent visits that were mainly due to Bombylius flies at low LI, (ii) high visitation rate,
prevailing due to butterflies, at intermediate LI and (iii) increasing visits of florivores at higher LI.

**Linking ecological heterogeneity and PBI patterns (question 2)**

The shifts in the visitation rate of mutualistic and antagonist insects influenced strength and patterns of PBIs over the LI gradient. Indeed, the individual patterns of flower accumulation showed a lower visitation rate at low LI. Instead, a prevalence of mutualistic (i.e. pollination) and antagonist (i.e. herbivory) interactions occurred at intermediate and high LI, respectively. Because the interplay between mutualistic and antagonistic interactions can generate selection conflicts on plant traits involved in insect attraction (Gómez 2003, 2008), such patterns of insect preferences could have affected the relationships between floral display size and pollinators. Indeed, although larger plant patches occurred at higher LI, here herbivory overcame pollination (Fig. 4). The frequency of herbivores at high LI, and their aptitude to perform multiple within-plant visits may have constrained pollination in many-flowered plants by reducing the available flowers, and by limiting the plant ability in attracting pollinators. Accordingly, herbivory was often found to alter plant traits involved in pollinator attraction, like scent (Lucas-Barbosa et al. 2011, 2016) and floral display (Penet et al. 2009; Söber et al. 2010). Such findings agree with a non-additive pattern between positive and negative interactions, highlighting that a large floral display size can become disadvantageous in antagonist-rich contexts (Gómez 2003, 2005).

**Table 4. Influence of light intensity and pollination treatment on seed germination and seedling survival.**

| Source of variation | Dependent variable | Wald chi² | P    |
|---------------------|--------------------|-----------|------|
| Light intensity     | Seed germination   | 13.860    | 0.008|
|                     | Seedling survival  | 17.487    | 0.002|
| Pollination treatment| Seed germination  | 11.756    | 0.003|
|                     | Seedling survival  | 17.637    | <0.001|
| Light intensity × Pollination treatment | Seed germination | 1.076 | 0.998 |
|                     | Seedling survival  | 1.551    | 0.992|

non-manipulated and manipulated flowers of treated plants (Knight et al. 2006; Wesselingh 2007). In our case, the seed-set of free-pollinated flowers from non-manipulated and manipulated plants was comparable. This suggested that: (i) the effect of within-plant reallocation processes promoted by pollen supplementation was negligible, and (ii) the presence of bags on experimental plants did not influence pollinators.

Seed-set comparison between naturally and experimentally pollinated flowers suggested a significant pollen limitation in the studied population. Pollen limitation is a frequent consequence of human disturbance (Eckert et al. 2010; Brys and Jacquemyn 2012), but in our case it seemed linked to the variations of insect activity occurred over the studied ecological gradient. In addition, as common in Dianthus (Jennersten 1987; Collin and Shykoff 2003; Gargano et al. 2009), D. balbisii showed a little ability for autonomous self-fertilization, and this can have favored the rise of pollen limitation. Indeed, functional limitation for self-fertilization limit the chances of reproductive assurance (Kalisz et al. 2004). Anyway, the extent of pollen limitation varied greatly in our system, being higher at the gradient extremes. Variations in insect assemblages were demonstrated to generate spatial patterns of pollen limitation (Gómez et al. 2010). Accordingly, our data evidenced that spatial structuring of pollen limitation can occur at very small scale, following the ecological gradients typical of transition habitats. Although pollinators are essential to allow pollen transfer (Ashman et al. 2004), herbivory can also influence pollen delivery rates. Florivory can reduce the pollen available to pollinators (Hargreaves et al. 2009; Harder and Aizen 2010), or limit the efficiency of plant–pollinator interactions (Penet et al. 2009; Söber et al. 2010; Lucas-Barbosa et al. 2011, 2016). Accordingly, the variations of both pollinators and herbivores contributed to model patterns of pollen limitation over the investigated gradient.

Pollen limitation also involves qualitative components of plant reproduction (i.e. offspring viability), which depend on the features of transferred pollen (Wilcock and Neiland 2002; Amat et al. 2011; Arceo-Gomez and Ashman 2014). The interplay between the structuring of plant populations and insect assemblages can affect the quality of pollen delivery (i.e. cross- vs. self-pollen) and, then, the viability in the subsequent offspring (Zorn-Arnold and Howe 2007; Le Cadre et al. 2008). In our study system, viability differed between progenies from hand-crosspollination and free pollination at low and high LI (Fig. 5B, C). Although Dianthus species are fully self-compatible, the presence of protandry and male-sterility promotes high outcrossing rates in their populations (Collin and Shykoff 2003). The consequent genetic load can result in severe inbreeding depression under enhanced selfing rates (Collin et al. 2006; Söber et al. 2010).
2009; Gargano et al. 2009, 2011, 2015). Therefore, the observed fitness variations can result from local variations in relative amount of cross- and self-pollinations. At intermediate $LI$, the fitness of naturally produced offspring was roughly equidistant from forced crossing and forced selfing (Fig. 5B, C). This would be compatible with the mixture of self- and cross-pollination typically occurring in plant populations (Wilcock and Neiland 2002). Instead, toward the gradient extremes the fitness of progenies from self-pollinated flowers declined, approaching the values detected after forced selfing (Fig. 5B, C). This suggested increased selfing rates. The accumulation of flowers in hermaphrodite phase on individuals, and the multiple within-plant visits due to Bombylius could have promoted inbreeding at low $LI$. Indeed, the presence of many open flowers on a single plant diminishes the effectiveness of dichogamy in preventing self-fertilization via geitonogamy (Pettersson 1992; Hidalgo and Hubera 2001). Such a limitation can be exacerbated if the prevalent pollinators travel over short-distance (Zorn-Arnold and Howe 2007). Instead, herbivory could have enhanced selfing rates at high $LI$. Indeed, by reducing flower display and plant attractiveness, massive florivory can increases the chances of self-fertilization due to sporadic pollinator visits (Penet et al. 2009; Söber et al. 2010). Then, the observed patterns of offspring viability can have reflected varying rates of inbreeding promoted by the interplay of: (i) habitat heterogeneity, (ii) insect fauna features (i.e. abundance, composition, and behaviour), and (iii) modality of pollen transfer.

Conclusions
Our work evidences how the ecological heterogeneity found at forest edges can generate small-scale patterns of relative insect abundance that are the premise for local variations in PBIs and plant fitness. Shifts in both mutualistic and antagonist insects can influence PBIs across such habitat gradients, contributing to generate quantitative and qualitative patterns of plant fitness. Such patterns reveal that in edge habitats plants can rely on a very small ecological space to optimize their reproduction. Therefore, the conservation of edge-dependent species requires the preservation of the habitat complexity typical of such transition contexts. Further work is needed to evaluate the effect on PBIs of temporal habitat dynamics involving these peculiar ecological contexts.

Supporting Information
The following additional information is available in the online version of this article—

Table S1. List and frequency of insects visiting flowers of *D. balbisii* during the observation trials, and their putative role in the context of the biotic interactions involving the study plant.

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Contributions by the Authors
D.G. originally formulated the idea, and developed methodology; D.G. and L.B. conducted fieldwork, D.G. and G.F. performed statistical analyses; D.G., G.F. and L.B. wrote the manuscript.

Conflict of Interest Statement
None declared.

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Literature Cited
Aguilar R, Ashworth L, Galetto L, Aizen MA. 2006. Plant reproductive susceptibility to habitat fragmentation: review and synthesis through a meta-analysis. *Ecology Letters* 9:968–980.
Amat ME, Vargas P, Gomez JM. 2011. Pollen quality limitation in the Iberian critically-endangered genus *Pseudomisopates* (Antirrhineae). *Plant Ecology* 212:1069–1078.
Arceo-Gomez G, Ashman TL. 2014. Local co-flowering community diversity influences pollen receipt, female reproductive success and mediates selection on flower longevity in *Mimulus guttatus*. *The American Naturalist* 183:E50–E63.
Ashman T-L, Knight TM, Steets JA, Amarasekare P, Burd M, Campbell R, Dudash MR, Mazer SJ, Johnston M, Mitchell RJ, Morgan MT, Wilson WG. 2004. Pollen limitation of plant reproduction: ecological and evolutionary causes and consequences. *Ecology* 85:2408–2421.
Ashman T-L, Penet L. 2007. Direct and indirect effects of a sex-biased antagonist on male and female fertility: consequences for reproductive trait evolution in a gender-dimorphic plant. *The American Naturalist* 169:595–608.
Bhattacharya M, Primack RB, Gerwein J. 2003. Are roads and railroads barriers to bumblebee movement in a temperate suburban area? *Biological Conservation* 109:37–45.
Block D, Werdenberg N, Erhardt A. 2006. Pollination crisis in the butterfly-pollinated wild carnation Dianthus carthusianorum. New Phytologist 169:699–706.

Brys R, Jacquemyn H. 2012. Effects of human-mediated pollinator impoverishment on floral traits and mating patterns in a short-lived herb: an experimental approach. Functional Ecology 26: 189–197.

Cariveau D, Irwin RE, Brody AK, Garcia-Mayeya LS, von der Ahe A. 2004. Direct and indirect effects of pollinators and seed predators to selection on plant and flower traits. Oikos 104:15–26.

Caruso CM, Peterson B, Ridley C. 2003. Natural selection on floral traits of Lobelia (Lobeliaceae): spatial and temporal variation. American Journal of Botany 90:1333–1340.

Chacón P, Armesto JJ. 2006. Do carbon-based defences reduce foliar damage? Habitat-related effects on tree seedling performance in a temperate rainforest of Chiloé Island, Chile. Oecologia 146:555–565.

Chalcoff VR, Aizen MA, Ezcurra C. 2012. Erosion of a pollination mosaic. The ecology of landscapes and re-...
Lennartsson Y. 2002. Extinction thresholds and disrupted plant-pollinator interaction in fragmented plant populations. Ecology 83:3060–3072.

López-Barrera F, Armesto JJ, Williams-Linera G. 2007. Fragmentation and edge effect on plant-animal interactions, ecological processes and biodiversity. In: Newton AC, ed. Biodiversity loss and conservation in fragmented forest landscapes: the forests of montane Mexico and temperate South America. Egham: CABI, 69–101.

Lucas-Barbosa D. 2016. Integrating studies on plant-pollinator and plant-herbivore interactions. Trends in Plant Science 21:125–133.

Lucas-Barbosa D, van Loon JJA, Dicke M. 2011. The effects of herbivore-induced plant volatiles on interactions between plants and flower-visiting insects. Phytochemistry 72:1647–1654.

Lucas-Barbosa D, Sun P, Hakman A, van Beek TA, van Loon JJA, Dicke M. 2016. Visual and odour cues: plant responses to pollination and herbivory affect the behavior of flowers visitors. Functional Ecology 30:431–441.

McCall AC, Irwin RE. 2006. Florivory: the intersection of pollination and herbivory. Ecology Letters 9:1351–1365.

Ollerton J, Winfree R, Tarrant S. 2011. How many flowering plants are pollinated by animals? Oikos 120:321–326.

Pauw A, Hawkins JA. 2011. Reconstruction of historical pollination rates reveals linked declines of pollinators and plants. Oikos 120:344–349.

Penet L, Collin CL, Ashman T-L. 2009. Florivory increases selfing: an experimental study in the wild strawberry, Fragaria virginiana. Plant Biology 11:38–45.

Peruzzi L, Gargano D. 2006. Dianthus ferrugineus Mill. vs. D. gulae Janka: nomenclatural considerations on the Italian yellow carnation. Taxon 55:781–784.

Petterson MW. 1992. Advantages of being a specialist female in gynodioecious Silene vulgaris s.l. ( Caryophyllaceae). American Journal of Botany 79:1389–1395.

Pignatti S. 1982. Flora d’Italia 1. Bologna: Edagricole.

Ries L, Fletcher RJ Jr, Battin J, Sisk TD. 2004. Ecological responses to habitat edges: mechanisms, models, and variability explained. Annual Review of Ecology, Evolution, and Systematics 35:491–522.

Risser PG. 1995. The status of the science examining ecotones. BioScience 45:318–325.

Sarlov-Herlin I. 2001. Approaches to forest edges as dynamics structures and functional concepts. Landscape Research 26:27–43.

Schoonhoven LM, van Loon JJA, Dicke M. 2005. Insect-plant biology. Oxford: Oxford University Press.

Söber V, Moora M, Teder T. 2010. Florivores decrease pollinator visitation in a self-incompatible plant. Basic and Applied Ecology 11:669–675.

Soper Gorden NL, Adler LS. 2016. Florivory shapes both leaf and floral interactions. Ecosphere 7:e01326.

Totland Ø. 2001. Environment-dependent pollen limitation and selection on floral traits in an alpine species. Ecology 82:2233–2244.

Verboven HAF, Aertsen W, Brys R, Hermy M. 2014. Pollination and seed set of an obligate outcrossing plant in a urban-peri-urban gradient. Perspectives in Plant Ecology, Evolution, and Systematics 16:121–131.

Wesselingh RA. 2007. Pollen limitation meets resource allocation: towards a comprehensive methodology. New Phytologist 174:26–37.

Wilcock C, Neiland R. 2002. Pollination failure in plants: why it happens and when it matters. Trends in Plant Science 7:270–277.

Zorn-Arnold B, Howe HF. 2007. Density and seed set in a self-compatible forb, Penstemon digitalis (Plantaginaceae), with multiple pollinators. American Journal of Botany 94:1594–1602.