Haemosporidian infections affect antioxidant defences in great tits *Parus major* but are not related to exposure to aerial pollutants

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SUMMARY

Haemosporidian parasites are widespread in avian species and modulate their ecology, behaviour, life-history and fitness. The prevalence of these vector-transmitted parasites varies with host intrinsic and extrinsic factors, such as host resistance and behaviour, and habitat-related characteristics. In this study, we evaluated the prevalence of avian haemosporidians in great tit *Parus major* populations inhabiting two areas with different degrees of exposure to aerial emissions from pulp mills, to assess if this type of pollution impacted parasite prevalence. We also compared the physiological condition of infected and uninfected individuals. Haemosporidian infection prevalence was high (58%), varied seasonally, but was not associated with air pollution exposure. Fledged birds presented higher infection rates than nestlings and infected fledged birds had higher levels of blood glutathione peroxidase activity. These results allow us to infer that infection by blood parasites may activate antioxidant defences, possibly to protect the organism from the negative oxidative stress side-effects of immune activation against parasites. Because oxidative stress is one of the mechanisms responsible for ageing and senescence and may affect fitness, the relationship between parasitism and oxidative stress markers should be further investigated through studies that include experimental manipulation of infection in model organisms.

Key words: host–parasite interactions, haemosporidia, *Plasmodium* spp, glutathione peroxidase, oxidative stress, pulp mill, avian malaria.

INTRODUCTION

Environmental degradation and presence of pollutants are likely to negatively impact wildlife. Apart from their direct effect, pollutants may affect the organism susceptibility to infection through direct immunotoxic effects (by interfering with immune receptor binding and triggering inappropriate and/or inhibiting immune responses), or through an increase of the overall stress levels of the organism which may lead to immunosuppression (Dobson and Foufopoulos, 2001; Martin and Boruta, 2014). The knowledge of the effects of pollutants on avian immune function in the wild remains limited. Grasman et al. (1996) reported that fish-eating birds exposed to polyclorinated byphenyls showed inhibited T-cell mediated immunity, which might have been the cause for a high incidence of bacterial infections in these high trophic level birds. Also, great tits exposed to heavy metal pollution from a metallurgic smelter complex had a significantly lower humoral immune responsiveness than birds sampled in areas farther away from the from the pollution source, after immunostimulation by injection of sheep erythrocytes (Snoeijjs et al. 2004). Environmental pollutants such as trace metals, have been reported to be associated with prevalence of *Plasmodium* infection in house sparrows *Passer domesticus*, suggesting immunotoxic effects of these contaminants (Bichet et al. 2013).

In this study, we aim to evaluate the haemosporidian community infecting great tit *Parus major* populations inhabiting two pine woods with different degrees of exposure to pollutants in west Portugal. We assessed the effect of environmental pollution (aerial emissions from pulp mills) on the prevalence of haemosporidians in those great tit populations. We also assessed the effects of haemosporidian infection on great tit physiology, namely on haematological, biochemical and morphological parameters, and whether their infection status was related with their ectoparasite levels. Haemosporidians are ubiquitous microscopic parasitic protozoa in birds, infecting blood cells and tissues and transmitted by blood-sucking insect vectors of the order Diptera (Valkiūnas, 2005).

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Those include the genera *Plasmodium, Haemoproteus* and *Leucocytozoon*, whose prevalence varies among geographical areas. Several factors affect the prevalence of these haemosporidians by affecting the presence/abundance of their vectors (climate and weather conditions, altitude, habitat conditions, including presence of freshwater bodies) and the susceptibility of the hosts (behaviour, phase of their life cycle, immune defences and resistance to a specific parasite), density of the host, among others (Sol et al. 2000; Valkuñas, 2005; Ortego and Espada, 2007; Cellier-Holzem et al. 2010; Knowles et al. 2011; Isaksson et al. 2013).

Haemosporidians are costly to their hosts because they consume their host’s metabolites and haemoglobin. Their effects can impact the organism in such a way that they may decrease its survival, leading directly (e.g. when invading naïve populations) or indirectly to its death, by decreasing its body condition and health, making it more vulnerable to predators and/or other diseases. They can also affect other life history traits such as decrease the breeding success of hosts (Merino et al. 2000; Marzal et al. 2005; Valkuñas, 2005; Asghar et al. 2015). Reported effects of their infection on health status in the wild include decrease in body mass and condition (Figuerola et al. 1999; Garvin et al. 2006; Valkuñas et al. 2006; Norte et al. 2009), decrease in total plasma proteins (Norte et al. 2009), decrease in haemoglobin levels (Norte et al. 2009; Krams et al. 2013), increase in total white blood cell (WBC) counts (Figuerola et al. 1999; Norte et al. 2009), modification of the leucocyte profiles (Ots and Höräk, 1998; Fokidis et al. 2008; Norte et al. 2009), decrease in T-cell-mediated response to phytohemagglutinin (PHA response; Navarro et al. 2003) and increase in circulating immunoglobulin levels (Ots & Höräk, 1998). However, the magnitude of those effects depends on the degree of co-evolution between the parasite and host and is also linked to environmental conditions (Lindström et al. 2005).

Co-infections are common in nature [not only dipteran-borne (Norte et al. 2009; Spurgin et al. 2012), tick-borne (Václav et al. 2011), but also among parasites with different transmission modes (Telfer et al. 2010)], which may be related not only with micro-habitat differences among individuals and degrees of exposure, but also with the individual’s quality (ability to direct energy to immune defences and anti-parasitic activities), or because infection by one parasite increases susceptibility to other parasites. Because, at least during the feather growth period, feather mite loads have been found to negatively correlate with condition-dependent traits (feather quality and growth rates (Thompson et al. 1997; Pérez-Tris et al. 2002; Figuerola et al. 2003) we assessed if infection by haemosporidians had an effect on feather mite loads.

Moreover, some haemoparasites consume their hosts’ haemoglobin and lead to erythrocyte destruction causing anaemia (Campbell, 1994). Body condition and total plasma protein were evaluated as indicators of the host’s nutritional status (Brown, 1996). Glutathione peroxidase is an antioxidant enzyme whose activity may increase to cope with reactive oxygen metabolites that may be produced during immune response to infection, to prevent oxidative damage to tissues (Ames et al. 1993). Total and differential WBC counts were assessed as indicators of the status of the immune system (Campbell, 1994), and also stress levels of the birds (Gross and Siegel, 1983). All these parameters (except haematocrit) were previously reported to correlate with haemoparasite infection in great tits (Norte et al. 2009). We expect that in the impacted area (that exposed to pulp mills aerial emissions) haemosporidian prevalence will be higher, except if exposure to pollution decreases vector abundance (Kozlov et al. 2005). We also expect that infected birds, especially nestlings and first-year birds during infection relapses (e.g. during demanding periods, such as the reproductive season for un-experienced first-year birds), will show signs of infection resulting for instance in anaemia, lower body condition and nutritional status, change in the WBC profile, increased antioxidant defences and higher feather mite loads, when compared with non-infected birds.

**MATERIALS AND METHODS**

**Study sites and sampling**

Great tits were captured in two pine woods of the Portuguese central west coast, near Figueira da Foz: Mata do Urso and Mata de Quiaios. Mata do Urso (42°55′N, 7°85′W) is located in the vicinity of two pulp mill industrial complexes (<1 km and 3 km to NE), and was defined as ‘polluted’. Mata de Quiaios (40°21′N, 8°35′W) is located ca. 20 km to the north of Mata do Urso. As prevalent winds in this area are from NNW, Mata de Quiaios is not exposed to pollution from the pulp mill complexes, and was defined as ‘control’. These definitions of polluted and control, based on the distances to the pollution source (pulp mills), were latter supported by analysis of mercury levels in great tit feather samples living in those areas. The two study areas are very similar in terms of habitat conditions and their main difference resides on their exposure to anthropogenic influence (Norte et al. 2010).

The pulp mills of the study area produce bleached Kraft pulp using an elemental chlorine-free method. Their gaseous effluents contain mostly carbon and sulphur oxides, organic and reduced sulphur compounds, nitrogen oxides and particles (StoraEnso, 2004, 2005), but may also contain volatile.
organochlorine compounds (Juuti et al. 1996), polycyclic aromatic hydrocarbons (PAH) and heavy metals from combustion of organic matter or fuel.

Birds were captured with mist nets from October 2004 to July 2005. During the breeding season both adults (when feeding 6–10 days old nestlings, hatching day = day 0) and nestlings (with 14 days old) were sampled in the nest boxes. Each bird captured in the mist nets was sexed and aged (1 year or adult) according to the plumage. Basic morphological measurements [body mass (g) and minimum tarsus length (mm)] were taken for all birds simultaneously to blood collection. Body condition index was obtained as the residuals of the linear regression of body mass on tarsus length (adults: \( b = 2.25 \pm 3.40 \), \( F_{1,23} = 17.80, P < 0.0001 \); nestlings: \( b = -0.38 \pm 4.58, F_{1,53} = 12.43, P < 0.0009 \)).

Wing flight feather mite loads were quantified according to the method of Behnke et al. (1999) in fledged birds. We visually inspected both primary, secondary and tertiary flight feather of the wing of great tits using a score system from 1 to 3 for each feather (0 = 0 mites; 1 = 1–10 mites; 2 = 11–30 mites; 3 = >30 mites). The blood sample was collected from the brachial vein, and it was used to make a thin-film smear and the remaining blood was separated into plasma and erythrocytes fractions and frozen at −80 °C for up to 45 days until analyses.

**Screening of blood parasites**

Blood smears were air dried and stained using the May–Grünwald–Giemsa method and scanned under 1000× magnification for WBC and heterophil/lymphocyte ratio (H/L) assessments. WBC count stands for the estimation of number of WBCs in approximately 10 000 erythrocytes. The H/L was measured by examining 50 WBCs because the repeatability of measurements on 50 and 100 WBCs was very high \( (r = 0.94 \pm 0.01, F_{1,109} = 31.86, P < 0.001) \).

Blood smears were screened for 10–15 min at low magnification \( (\times 400) \) to detect *Leucocytozoon* sp., *Trypanosoma* sp. and microfilaria. For the detection of *Plasmodium* and *Haemoproteus*, 200 fields were observed under oil immersion (at 1000× magnification) using Olympus Microscope CX31 (Olympus Corporation, Tokyo, Japan; Valkiūnas, 2005).

**Physiological markers’ assays**

Total plasma protein (protein – mg mL\(^{-1}\)) was measured using the Bradford Protein Assay, based on the Bradford dye-binding procedure (Bradford, 1976). For glutathione–peroxidase measurements (GPx; \( \mu M \min ^{-1} g^{-1} \) haemoglobin) we used the method of Paglia and Valentine (1967) as modified by Silva and Santos (1991), which consisted of measuring the rate of oxidized glutathione (GSSG) formation at 340 nM as NADPH is converted to NADP+. Enzymatic activity and total protein were measured in a Sunrise Inst. Microplate reader. Hematocrit (Hct) was measured as the percentage of the length of the part of the capillary tube occupied by erythrocytes in relation to the total length of the capillary tube occupied by blood, after its centrifugation for 10 min at 1032 g. Erythrocyte haemoglobin index (Hb; g L\(^{-1}\)) was defined as the haemoglobin content in the 4× diluted haemolsate of erythrocytes and was not equivalent to whole blood haemoglobin concentration or dependent on hematocrit. It was measured with the cyanmethaemoglobin method at a wavelength of 540 nm (van Kampen and Zijlstra, 1961) using a commercial kit (BioSystems S.A., Barcelona, Spain). Haemoglobin was measured in addition to Hct because it may be a more sensitive and long-lasting indicator of anaemia (O’Brien et al. 2001; Norte et al. 2009).

**Statistical analysis**

Recaptured birds within 60 days of the first capture were excluded from the analyses; therefore, birds captured within the same season were not included, which gives time for the birds to recover of any effects of blood sampling on physiological parameters of interest. Only eight birds (8/138) were captured twice over different seasons and were included in the analyses. We evaluated if morphological and physiological parameters and feather mite loads were correlated with each other in fledged and nestling birds separately using Spearman correlations (Table S1). We compared prevalence of infection between nestling and fledged birds with a Mann–Whitney test using the mean prevalence of infection per nest. Because infection prevalence was significantly different among these groups (see below), further analyses were performed separately. Also, because nestlings are actively growing, their physiology is different (Kostelecka-Myrcha et al. 1973), and this should be taken into account when analysing and interpreting the data.

We evaluated how infection by haemosporidians in fledged birds was affected by intrinsic (age, sex) and extrinsic factors (season, pollution exposure) using Generalized Linear Models, with binomial distributions and logit link function. We used a stepwise approach, by including all explanatory variables in the models at first and removing them one by one, until the model contained only significant explanatory variables. Interactions between significant variables were also tested for significance and model improvement. Only the extrinsic effect of pollution exposure was tested on nestling infection status using a Mann–Whitney test using the mean prevalence per nest. The effects of infection by haemosporidians on the physiology of both fledged birds and nestlings was tested separately using General
Correlates of infection in nestling birds

No effects of being raised in the area exposed to aerial pollutants from pulp mills was detected on prevalence of haemosporidian infection ($Z_{1,11} = 0.65, P = 0.51$). No effects of infection by haemoparasites were detected on any of the physiological parameters evaluated (all $P > 0.05$).

Correlates of infection in fledged birds

Haemosporidian infection in fledged birds was affected by season (prevalence autumn = 25/37; winter = 20/23; spring = 7/15; estimate [spring]
−1.16 ± 0.37; estimate [winter] = 1.09 ± 0.41, $\chi^2_{2.79} = 11.69, P = 0.003$) with higher infection prevalence in winter and lower in spring. Haemosporidian infection affected GPx activity of fledged birds with infected birds having higher GPx antioxidant activity levels (Table 2, Fig. 3). In the same model, GPx was also affected by the age of the bird, and exposure to aerial pollution (with higher values for older birds, and in the study area near pulp mills; Table 2). Total feather mite scores varied between 0 and 50, but did not differ significantly between Haemospora infected and non-infected birds (mean ± s.e. infected = 7.5 ± 1.33; non-infected: 6.3 ± 1.66). Primary feather and secondary and tertiary feather mite scores also did not differ significantly between infected and non-infected birds. None of the other measured morphological and physiological parameters were affected by infection status ($P > 0.05$).

**DISCUSSION**

Our results do not support the hypothesis that haemosporidian infection is related with anthropogenic pressure, in the form of exposure to pulp mill aerial emissions. Anthropogenic habitat degradation may alter habitat condition for parasites, vectors and hosts, thereby altering prevalence and intensity of parasitism. Prevalence of haemoparasites in impacted areas may be lower than in conserved areas, but this depends on study areas/habitat types and parasite lineages (Chasar *et al.* 2009; Evans *et al.* 2009). In the specific case of the impact of pollutants, they can have direct immunotoxic effects on the host by inhibiting its immune system or contribute to the stress of the organism and increase its corticosterone levels, and its ability to cope with parasites (Martin and Boruta, 2014). Air emissions from pulp mills result mainly from the combustion of fuels to produce energy for pulp making processes. The gaseous effluents contain carbon dioxide and sulphur dioxide, and other sulphur oxides, reduced sulphur and organic sulphur compounds, nitrogen oxides (NOx), carbon monoxide and particles [26, 27]. They may also contain volatile organochlorine compounds [28], PAH and metals from the combustion of organic matter or fuel. In a previous study (Norte *et al.* 2010) found that

**Fig. 1.** *Plasmodium* spp. observed in May-Grünwald-Giemsa-stained blood smears from great tits (*Parus major*). (A, B) trophozoites; (C, D) erythrocytic meronts; (E) macrogametocyte and, (F) microgametocyte. Arrows indicate infected erythrocytes.
mercury concentration in great tit feather samples from Mata do Urso, near the pulp mill complexes, was significantly higher than the nearby unpolluted control area, Mata de Quiaios. Trace metals were shown to depress the immune system in birds (Snoeij et al. 2004) and mercury, in particular, was reported to depress the immune system (Kenow et al. 2007; Hawley et al. 2009), both in captivity and in the wild. However, in this study, we found no effects of pollution exposure on the prevalence of haemosporidians. Nonetheless, because we compared only two study areas, although chosen in order to be as similar as possible, we cannot exclude the hypothesis that they could have differed in other environmental characteristic, besides pollution exposure, that may have affected haemosporidian prevalence and confounded our results. We also found no evidence from the physiological profile of great tits (WBC or H/L) that they differed between polluted or control areas in their immune status.

Prevalence of haemosporidians is often reported to vary seasonally, which is related with the fluctuations in populations of vectors, but also

**Table 2. General Linear Model of the effects of haemosporidian infection and other co-variables (exposure to pollution and age) on glutathione-peroxidase activity of great tits Parus major**

| GPx | $R^2 = 0.22; F = 6.63; P < 0.0005; n = 73$ |
|-----|------------------------------------------|
|     | Estimate ± s.e. | $t$  | $P$   |
| Intercept                      | 11.86 ± 0.27 | 43.99 | < 0.0001 |
| Haemosporidian infection [infected] | 0.60 ± 0.27 | 2.24  | 0.029   |
| Pollution exposure [polluted]   | 0.58 ± 0.25  | 2.31  | 0.024   |
| Age [adult]                    | 0.70 ± 0.25  | 2.75  | 0.008   |

Final models include only significant variables.
susceptibility of the hosts, associated with energy demanding periods of their life cycles. Haemosporidian prevalence usually increases in spring in temperate regions (Hatchwell et al. 2000; Norte et al. 2009), and an autumn peak may also be observed (Cosgrove et al. 2008). In this study, we found that haemosporidian prevalence had a different pattern, being lower in spring, intermediate in autumn and higher in winter. Ventim et al. (2012) also found that Plasmodium infections in Cetia cetti and P. domesticus increased from spring, to autumn and winter in Portuguese reedbeds, which could be related with lower food availability in those habitats during those periods compared with spring, making the birds more vulnerable to infections. This may also occur in pine woods, a relatively poor habitat, compared with deciduous and mixed woods, in which the majority of European studies were performed.

Interestingly we found that fledged birds infected with haemosporidians had higher levels of glutathione peroxidase. This effect was significant even after controlling for the effects of pollution exposure and age, which also increased GPx antioxidant activity. This may be explained by the activation of the immune system in response to parasite infection through cell-mediated immune response and inflammation mechanisms (Isaksson et al. 2013). This increases the levels of reactive oxygen species, which are released by heterophils to attack the parasite. Because this response in not specific and does not distinguish between self and parasite tissue, it may cause oxidative damage to all types of adjacent molecules and can lead to haemolysis and cellular dysfunction, increasing the risk of oxidative stress to the organism (Isaksson et al. 2013). The increase of glutathione peroxidase activity may be one protective action to recover from the potential unbalance in oxidative status (Meister and Anderson, 1983). Our previous study, in a nearby mixed deciduous woodland, also showed that both Plasmodium and Leucocytozoon infected great tits had significantly higher glutathione peroxidase activity (Norte et al. 2009). Isaksson et al. (2013) also reported that the levels of total and oxidized glutathione in great tits were significantly affected by the Plasmodium infection status, which could be related with the activity of glutathione peroxidase (Meister and Anderson, 1983). Moreover, a meta-analysis has shown that oxidative stress may indeed be a physiological cost of immune activation because the activation of the immune response changes oxidative stress markers, including glutathione peroxidase activity, although the direction of the effect may not always be consistent (Costantini and Møller, 2009). Therefore, it is important to measure a set of oxidative stress markers to have a better picture of the oxidative status of the organism because different protective mechanisms may be activated at different stages (Costantini and Møller, 2009). Although haemosporidian infection may have elicited an immune response and inflammation in infected birds, this was not reflected in the total and differential leukocyte counts, although Plasmodium, Haemoproteus and Leucocytozoon have been shown to cause leucocytosis (Figuerola et al. 1999). Glutathione peroxidase activity was also affected by pollution exposure, with higher values in birds inhabiting the vicinity of pulp mills. This confirms previous results that great tits exposed to
aerial emissions from pulp mills show upregulated antioxidant defence mechanisms (Norte et al. 2010).

We did not find effects of haemosporidian infection in haemoglobin or haematocrit levels, which is in accordance with Dufva (1996). A literature review showed that results have been contradictory concerning the relationships between haematocrit and haematozoan parasite infection (Fair et al. 2007), but haemoglobin levels have been reported to be more sensitive indicators of parasite induced anaemia (Norte et al. 2009). Also birds suffering from anaemia may move less and be less often captured making it difficult to detect such relationships in wildlife studies. Possibly, the birds in our study were in the chronic phase of infection, as suggested by the low parasitaemia intensity observed in their blood smears, and anaemia is noticeable mainly during the acute phase of infection (Cornet et al. 2013).

We did not find an association among haemosporidians and feather mite infestation loads. Positive associations between blood parasites and other parasite types, with different transmission modes, especially ectoparasites, have not been commonly found in nature. Kleindorfer et al. (2006) and Norte et al. (2013) found no association between blood parasite infection and tick infestation loads, possibly due to different branches of the immune system acting against different parasites; also, these associations may only be detectable at high infection rates. But Telfer et al. (2010) revealed that infections by cowpox virus (directly transmitted), Babesia microti (tick-transmitted), Bartonella spp. (flea-transmitted) and Anaplasma phagocytophilum (tick-transmitted) were connected in field voles (Microtus agrestis), supporting the increased susceptibility hypothesis. Although, in our study, primary feather mite loads were negatively correlated with body condition, suggesting that they could be indicators of the individuals’ quality, we did not find an association among them and haemosporidian infection. The relatively small sample size of our study might have limited the detection of other potential relationships between infection and physiological markers or the detection of other environmental factors that might affect infection status. Also, because different haemosporidian genera have different ecologies and habitat/vector requirements, the hemoparasite genera identification through molecular methods should help to detect environmental relationships.

The prevalence of haemosporidian infection reported in this study (58%) was higher than that reported for fledged great tits inhabiting a mixed wood in Portugal situated approximately 40 km away (31-2%; Norte et al. 2009). In that population, Plasmodium was also the most common Haemoproteus, just as in the study by Antunes (2007) in passerines in central Portugal, in which the most prevalent haemoparasite was Plasmodium, followed by Leucocytozoon and by Haemoproteus. Variation in the prevalence of haemosporidian parasites is expected, even at small spatial scales, and related with habitat and host characteristics, vector abundance and interactions among these variables (Shurulinkov and Chakarov, 2006; Knowles et al. 2011; Isaksson et al. 2013). We detected no Trypanosoma infection in this study, which agrees with the lower prevalence of Trypanosoma in great tits previously reported for Central Portugal (1%; Norte et al. 2009). Moreover, to our knowledge there are also no reports of microfilaria in Portugal from avian blood smears (Tomás, 2014). The fact that this study was based on the microscopic evaluation of blood smears might have underestimated the prevalence of blood parasites, especially in low-intensity infections and/or during their chronic phase, when there are no gametocytes in circulation. However, similar prevalences of malarial and Leucocytozoon parasites may be obtained by microscopy and PCR-based techniques in naturally infected birds if good-quality smears are examined by skilled investigators (Valkiūnas et al. 2008). The fact that fledged birds presented a higher prevalence of infection than nestling birds was expected given the higher probability of previous contact with infected vectors (Wood et al. 2007). This was shown for house martins (Delichon urbica) as the probability of blood parasite infection increased as individuals age (Marzal et al. 2015), which, may also be explained by immunosenescence (Knowles et al. 2011). However, a decrease in prevalence of blood parasites with age has also been reported, depending on the genus/lineage of the parasite. Sol et al. (2003) showed that Haemoproteus infections declined as hosts age, which may be explained by acquired immunity (Knowles et al. 2011). Although the prepatent period varies with haemosporidian genera and species (from 2 days to several months (Valkiūnas, 2005), we detected infected nestlings when they were only 15 days old, meaning that they were infected in their first days of life. Prevalent malarial parasites and Leucocytozoon (36–67%) in 5–15 days old nestlings have been previously reported (Fargallo and Merino, 2004; Zehtindijev et al. 2012; Calero-Riestra and Garcia, 2016).

This study contributed to a better knowledge of the correlates of haemosporidian infection in wild birds in southwestern Europe. There was no evidence that anthropogenic pressure impacted the prevalence of haemosporidians in this common passerine species, the great tit, but we reported that parasite infection, even at low intensities, affects antioxidant defences of the birds. This result is of even greater significance because it is supported by previous evidence from another great tit population in a nearby study area of a different habitat type (Norte et al. 2009). Our results also show that great
Haemopirridians affect glutathione peroxidase in the great tits exposed to aerial emissions from pulp mills may incur in increased physiological harm when infected because both pollution exposure and haemosporidian infection affect their oxidative balance, with potential increased detrimental impacts on their fitness. Further experimental studies manipulating haemosporidian infection and evaluating a wider range of oxidative stress markers will help to further elucidate the mechanisms of protection against these ubiquitous parasites.

SUPPLEMENTARY MATERIAL
The supplementary material for this article can be found at https://doi.org/10.1017/pao.2017.4

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CONFLICT OF INTEREST
None.

ETHICAL STANDARDS
Bird capturing and sample collection complied with ethical standards and was performed with permission from national authorities.

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