Microbial communities in peatlands along a chronosequence on the Sanjiang Plain, China

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Microbial communities play crucial roles in the global carbon cycle, particularly in peatland ecosystems under climate change. The peatlands of the Sanjiang Plain could be highly vulnerable to global warming because they are mainly located at the southern limit of northern peatlands. In this study, the alpha diversity and composition of bacterial communities in three different minerotrophic fens along a chronosequence were investigated. We captured a rich microbial community that included many rare operational taxonomic units (OTUs) but was dominated by a few bacterial classes that have frequently been detected in other peatland ecosystems. Notably, a large diversity of methanotrophs affiliated with Alpha- and Gammaproteobacteria was also detected. Bacterial alpha diversity and composition varied as a function of peat depth and its associated physical-chemical properties, such as total carbon, total nitrogen, pH and bulk density. We also found that bacterial community turnover (beta diversity) to be significantly correlated with soil age, whereas bacterial alpha diversity was not.

Despite covering only 6–8% of the world’s terrestrial ecosystems, northern peatlands store approximately 30% of the global soil carbon (C)1. Thus, as one of the largest atmospheric carbon sinks and CH4 sources, northern peatlands have played an important role in global C cycles and climate change throughout the Holocene (from approximately 11600 years ago to the present)2–5. In peatlands, the sequestration of C arises from an imbalance between inputs via primary productivity and losses through microbial decomposition, which releases C to the atmosphere and as dissolved organic carbon (DOC) exports4,5. Both environmental and physico-chemical conditions impact C cycling through primary production and decomposition processes. Methane production is a consequence of the permanently waterlogged condition of the peat layers, where anaerobic decomposition prevails, and many studies have shown that microorganisms play a crucial role in C decomposition and methane production process6–10. Although microbial communities in the peatlands of Europe, America, Canada and the UK have been investigated11–15. However, there are relatively few studies on the peatlands in more temperate regions14–16. The Sanjiang Plain, located in the temperate climate region, is the largest area of freshwater marshlands in China17. As these peatlands are mainly located at the southern limit of northern peatlands, they may be more vulnerable to global warming than other areas18. Thus, it is imperative that we deepen our understanding of the microorganisms in this ecosystem.

The peatlands of the Sanjiang Plain have developed under certain topographic conditions during the Holocene or earlier19, and they vary in age from 500 to 10,000 years20. Correspondingly, these areas differ with regard to C accumulation rate. Accordingly, this chronosequence of peatlands offers an exceptional opportunity for studying links between the chronological characterization (soil age), soil properties and microbial community structure in peatland ecosystems. The influences of vegetative communities, water hydrology and soil properties in peatland microbial communities have been examined in many studies21–24, though the association between chronological characterization and the distribution of the microbial community structure has been neglected to date.

In the present study, high-throughput Illumina sequencing of 16S rRNA genes was applied to examine the bacterial communities of the Sanjiang Plain, the southern edge of northern peatlands. Peat cores were collected from three fens that began to develop in this area during different periods. The greater sequencing depth achieved by high-throughput sequencing allows for the capture of the less abundant and uncultured taxa and thus supplies

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a more thorough characterization of peatland bacterial diversity. Moreover, such chronological characterization facilitates the description of potential links between bacterial communities and soil age as well as the C accumulation rates.

**Results**

**Physico-chemical and chronological characterization of peat cores.** Nine peat cores were retrieved from fens, S (Shenjiadian), H (Honghe), and Q (Qindelie), on the Sanjiang Plain, northeastern China (Fig. 1). With a depth of 100 cm, the peatlands located in Q are shallower than those at the other sites; mud deposits occur at depths greater than 100 cm. Accelerate mass spectrometry (AMS) dating results indicate that Sanjiang Plain peatlands began to develop over a wide range of dates during the Holocene (Fig. S1). The peat was dated from 637 to 11,496 cal. yr BP (Table 1). Peat cores from Q exhibited the most ancient chronological dates, with even the 30–60-cm section being dated to 3136 ± 1325 cal. yr BP, which was older than the age of the bottom section (100–200 cm) of the youngest fen, S at 2881 ± 1339.5 cal. yr BP. Overall, the mean C accumulation rates of the peat cores decreased with depths with one exception, the maxima of the mean C accumulation rates of the peat cores from S were observed in the 60–100-cm section (Table 1).

A summary of the soil physico-chemical characteristics is presented in Table 1. Soil pH was consistently acidic and varied from 4.80 to 5.91. Soil total C (TC) and N (TN) ranged from 29.74 to 458.54 g/kg and from 2.95 to 21.62 g/kg, respectively. Overall, they were decreased with depths at all sites, and the soils from S, the youngest fen, showed the highest total C and N contents. The C/N ratio was approximately 1.5 times lower in the lowest section than in the surface section of the same site.

Soil age (analysis of variance (ANOVA), F = 7.320, p = 0.001), TC (F = 4.333, p = 0.012) and pH (F = 9.297, p < 0.001) differed vertically. However, the C accumulation rate (F = 0.330, p = 0.804), TN (F = 7.320, p = 0.216), C/N ratio (F = 1.505, p = 0.234) and bulk density (F = 0.603, p = 0.618) did not. Interestingly, we found that TN (Pearson test, r = −0.360, p = 0.039), TC (r = −0.505, p = 0.003) and the C accumulation rate (r = −0.349, p = 0.012) were significantly correlated with soil age at 0.05 significance level.

**Figure 1.** Map of the study region showing locations of the sampling sites on the Sanjiang Plain. The map (including the inset map of China) was generated by Zhenqing Zhang using ArcGIS 10.0 (http://www.esrichina.com.cn/software/product/ArcGIS/).
0.046), were negatively correlated with peat age, whereas pH ($r = 0.541$, $p = 0.001$) was positively correlated with peat age.

### Bacterial species richness and equitability

Concentrations of DNA extracted from peat soils across different sites and depths were ranged from 88.2 to 240.9 ng/µL. After amplification of the V3-V4 region of the 16 rRNA, 1,789,830 sequences from 24 soil samples were sequenced using Illumina HiSeq (Tables 2), and 1,205,425 sequences remained after quality control. After resampling with 21,811 sequences, the number of operational taxonomic units (OTUs) per sample ranged from 357 to 1542. The Shannon, Simpson, Chao1, ACE and PieLou equitability indices were calculated to estimate microbial richness and equitability (Table 2). The Shannon (ANOVA, $F = 3.978$, $p = 0.017$) and PieLou equitability ($F = 4.329$, $p = 0.012$) indices changed significantly from the surface to the bottom layer of the peat cores. However, no change was observed for the OTU number ($F = 2.521$, $p = 0.077$) or Simpson ($F = 1.867$, $p = 0.157$), Chao1 ($F = 1.922$, $p = 0.148$), and ACE ($F = 1.670$, $p = 0.195$) indices.

### Microbial community composition

Diverse bacteria were found at different peat depths. OTUs were affiliated with 55 bacterial phyla, with the 10 most abundant phyla being Proteobacteria (54.29%), Actinobacteria (16.07%), Acidobacteria (10.10%), Bacteroidetes (6.06%), Firmicutes (5.57%), Chloroflexi (3.30%), Gemmatimonadetes (0.81%), Caldserica (0.57%), TM7 (0.51%) and AD3 (0.28%) (Fig. 2). The bacteria ranking in the top 10 in terms of abundance at the genus level belong to Acidobacteria, Proteobacteria and Actinobacteria (Fig. 3).

| Sample | Location     | Depth     | Total C (g·kg$^{-1}$) | Total N (g·kg$^{-1}$) | C/N ratio | pH     | Bulk density (mg·cm$^{-3}$) | AMS $^{14}$C age (14Cy rBP) | C accumulation rate (g C·m$^{-2}$ yr$^{-1}$) |
|--------|--------------|-----------|-----------------------|-----------------------|-----------|--------|----------------------------|-----------------------------|---------------------------------|
| S1     | Shenjiang1   | 0–30 cm   | 340.62                | 21.62                 | 15.75     | 5.24   | 0.588                      | 863                         | 101.37                          |
| S2     | Shenjiang2   | 30–60 cm  | 289.15                | 11.94                 | 24.38     | 5.38   | 0.535                      | 820                         | 84.85                           |
| S3     | Shenjiang3   | 60–100 cm | 316.33                | 11.62                 | 27.22     | 5.31   | 0.462                      | 637                         | 36.98                           |
| S4     | Shenjiang2   | 60–100 cm | 249.70                | 30.46                 | 11.48     | 5.36   | 0.172                      | 1554                        | 31.38                           |
| S5     | Shenjiang2   | 100–200 cm| 29.74                 | 11.88                 | 24.38     | 5.36   | 0.046                      | 1903                        | 13.92                           |
| S6     | Shenjiang3   | 60–100 cm | 297.44                | 17.26                 | 18.54     | 5.38   | 0.412                      | 3275                        | 15.55                           |
| S7     | Shenjiang3   | 100–200 cm| 269.36                | 14.51                 | 18.56     | 5.62   | 0.680                      | 3776                        | 21.89                           |
| S8     | Shenjiang3   | 100–200 cm| 336.08                | 15.01                 | 22.39     | 4.98   | 0.390                      | 1342                        | 45.39                           |
| H1     | Honghe1      | 0–30 cm   | 286.88                | 19.11                 | 15.01     | 4.91   | 0.577                      | 683                         | 70.92                           |
| H2     | Honghe2      | 30–60 cm  | 423.85                | 34.05                 | 12.45     | 5.44   | 0.574                      | 764                         | 205.66                          |
| H3     | Honghe3      | 60–100 cm | 68.95                 | 2.95                  | 23.37     | 5.32   | 1.181                      | 3741                        | 8.22                            |
| H4     | Honghe2      | 60–100 cm | 331.76                | 18.19                 | 18.23     | 5.56   | 0.428                      | 1412                        | 55.86                           |
| H5     | Honghe3      | 100–200 cm| 391.40                | 19.04                 | 20.56     | 5.65   | 0.407                      | 1029                        | 81.99                           |
| H6     | Honghe3      | 100–200 cm| 283.30                | 9.09                  | 31.16     | 5.59   | 0.379                      | 6275                        | 2.55                            |
| H7     | Honghe2      | 100–200 cm| 254.18                | 7.44                  | 34.16     | 5.57   | 0.382                      | 2510                        | 24.33                           |
| H8     | Honghe3      | 141.34                | 5.02                  | 28.16     | 5.51   | 0.973                      | 3727                        | 10.39                           |
| H9     | Honghe3      | 278.36                | 19.00                 | 14.65     | 5.60   | 0.382                      | 8166                        | 1.84                            |
| H10    | Honghe2      | 100–200 cm| 338.80                | 20.74                 | 16.34     | 5.60   | 0.380                      | 5053                        | 9.97                            |
| H11    | Honghe2      | 89.23                 | 8.59                  | 10.39     | 5.73   | 0.384                      | 5625                        | 6.65                            |
| Q1     | Qindelie1    | 0–30 cm   | 340.12                | 16.10                 | 26.72     | 4.80   | 0.335                      | 2085                        | 9.29                            |
| Q2     | Qindelie2    | 382.54                | 15.28                 | 25.04     | 5.14   | 0.448                      | 924                         | 56.74                           |
| Q3     | Qindelie3    | 378.24                | 15.57                 | 24.30     | 4.96   | 0.590                      | 1020                        | 61.72                           |
| Q4     | Qindelie1    | 267.59                | 13.39                 | 19.98     | 5.30   | 0.486                      | 4626                        | 14.66                           |
| Q5     | Qindelie2    | 458.54                | 16.79                 | 27.33     | 5.30   | 0.218                      | 2088                        | 12.99                           |
| Q6     | Qindelie3    | 156.00                | 13.04                 | 11.96     | 5.27   | 0.905                      | 2694                        | 0.86                            |
| Q7     | Qindelie3    | 141.91                | 12.13                 | 11.70     | 5.80   | 0.76                        | 11496                       | 0.99                            |
| Q8     | Qindelie2    | 181.95                | 14.08                 | 12.92     | 5.55   | 0.735                      | 4915                        | 5.44                            |
| Q9     | Qindelie3    | 44.72                 | 6.67                  | 6.7       | 5.91   | 0.812                      | 4410                        | 4.31                            |

Table 1. Soil properties, including the AMS dating result and carbon accumulation rate, of samples from 9 peat cores from the Sanjiang Plain. (The peatlands located in Qindelie were shallower than those at the other sites, with a depth of 100 cm).
## Community comparison across all sites and depths.

A Venn diagram was used to compare the similarities and differences among the communities in the different samples. Because they were more likely to reflect

### Table 2. Diversity indices of the bacterial communities of Shengjiadian (S), Honghe (H) and Qindelie (Q) fens.

| Sample name | Sequenced reads | OTU number | Shannon | Simpson | Chao1 | ACE | PieLou | equitability |
|-------------|-----------------|------------|---------|---------|-------|-----|--------|--------------|
| S1          | 25,518          | 1007       | 6.487   | 0.951   | 1194.4 | 1233.4 | 0.650 |
| S2          | 39,835          | 1542       | 8.102   | 0.99    | 2152.6 | 2219.1 | 0.765 |
| S3          | 41,599          | 1119       | 7.542   | 0.986   | 1472.4 | 1515.1 | 0.745 |
| H1          | 50,533          | 1000       | 7.643   | 0.988   | 1285.0 | 1289.6 | 0.767 |
| H2          | 45,427          | 1351       | 7.866   | 0.982   | 1549.1 | 1537.1 | 0.756 |
| H3          | 25,962          | 1215       | 7.603   | 0.979   | 1413.5 | 1430.8 | 0.742 |
| Q1          | 40,211          | 1006       | 6.856   | 0.97    | 1327.4 | 1371.6 | 0.687 |
| Q2          | 36,181          | 1490       | 7.978   | 0.983   | 1763.1 | 1721.2 | 0.757 |
| Q3          | 43,837          | 855        | 6.8     | 0.974   | 1066.3 | 1078.9 | 0.698 |
| S4          | 37,855          | 1069       | 7.145   | 0.981   | 1454.2 | 1504.3 | 0.710 |
| S5          | 37,750          | 869        | 6.705   | 0.972   | 1187.7 | 1205.4 | 0.687 |
| S6          | 26,248          | 933        | 7.056   | 0.976   | 1113.3 | 1162.3 | 0.715 |
| H4          | 44,583          | 973        | 6.425   | 0.92    | 1171.2 | 1176.9 | 0.647 |
| H5          | 22,157          | 1275       | 7.855   | 0.985   | 1500.8 | 1513.8 | 0.761 |
| H6          | 32,546          | 1216       | 7.502   | 0.978   | 1381.5 | 1474.8 | 0.732 |
| Q4          | 41,677          | 1090       | 6.755   | 0.965   | 1568.3 | 1571.9 | 0.669 |
| Q5          | 32,136          | 967        | 6.968   | 0.976   | 1130.4 | 1217.5 | 0.703 |
| Q6          | 44,614          | 357        | 4.498   | 0.828   | 447.8  | 492.1  | 0.530 |
| S7          | 25,647          | 1267       | 7.73    | 0.982   | 1468.6 | 1504.7 | 0.750 |
| S8          | 47,196          | 725        | 5.921   | 0.953   | 941.8  | 999.8  | 0.623 |
| S9          | 42,905          | 808        | 5.74    | 0.928   | 1174.9 | 1252.5 | 0.594 |
| H7          | 37,366          | 1217       | 7.265   | 0.975   | 1556.9 | 1614.0 | 0.709 |
| H8          | 38,994          | 933        | 6.482   | 0.969   | 1250.0 | 1284.9 | 0.657 |
| H9          | 21,811          | 1363       | 7.524   | 0.981   | 1860.5 | 1870.1 | 0.723 |
| S10         | 39,687          | 921        | 6.957   | 0.978   | 1245.1 | 1260.8 | 0.707 |
| S11         | 27,396          | 594        | 6.293   | 0.969   | 696.6  | 716.6  | 0.683 |
| S12         | 27,489          | 1208       | 7.809   | 0.98    | 1372.7 | 1399.0 | 0.763 |
| Q7          | 39,499          | 596        | 4.952   | 0.898   | 814.0  | 887.1  | 0.537 |
| Q8          | 31,287          | 833        | 6.942   | 0.979   | 973.0  | 1031.8 | 0.716 |
| Q9          | 45,560          | 704        | 5.24    | 0.92    | 919.9  | 988.9  | 0.554 |
| H10         | 35,835          | 821        | 6.145   | 0.954   | 1177.3 | 1149.6 | 0.635 |
| H11         | 38,421          | 1040       | 6.324   | 0.95    | 1336.2 | 1409.3 | 0.631 |
| H12         | 37,663          | 988        | 6.482   | 0.957   | 1307.3 | 1339.9 | 0.652 |

### Figure 2. Bacterial community structure variation in the Shengjiadian (S), Honghe (H) and Qindelie (Q) fens.

The relative abundance of bacteria at the phylum level is shown. Each bar represents the relative abundance of each sample. Each color represents a particular phylum. The numbers associated with the sample names indicate the sampling depth.
community function, only highly abundant OTUs (relative abundance higher than 0.1% and detected in more than 6 samples) were used for the calculations. The bacterial communities of the surface peat of the S, H, and Q fens shared 42 common OTUs (Fig. 4a), and bacterial communities varied with depth at each site. For example, in the soil sampled from S, the unique OTU numbers of the 0–30, 30–60, 60–100, and 100–200-cm sections were 56, 28, 18, and 25, respectively (Fig. 4b).

Bacterial communities differed across depths and sites. The relative abundance of the dominant phylum, Proteobacteria, increased with increasing depth across all sites and was higher in younger peatlands, in contrast, Actinobacteria and Acidobacteria sequences decreased with increasing depth (Fig. 2). From top to the bottom, the relative abundance of Actinobacteria in the peat cores from S, H and Q fens, decreased from 23.60% to 13.16%, from 15.10% to 12.62%, and from 24.24% to 8.06%, respectively, and that of Acidobacteria decreased from 16.29% to 1.53%, from 12.21% to 4.72%, and from 16.78% to 4.50%, respectively. Notably, a large diversity of methanotrophs affiliated with Alpha- and Gammaproteobacteria was detected (Fig. 5). Methylocystaceae were the dominant methanotrophs across all peat samples and were primarily observed in the deep and aged soils.

Environmental variation affects community structure. We used constrained correspondence analysis (CCA) to analyze variations in bacterial community structure and its associations with environmental variables. Overall, the measured environmental variables were not significantly correlated with bacterial community structure (Mantel test, r = 0.07, p = 0.239), and only those environmental variables that showed a significant correlation (p < 0.05) with the bacterial community were plotted as vectors (Fig. 6). On the horizontal axis (CCA1, 30.41% of constrained variability), the most influential constraining variable was pH (biplot score = 0.84), followed by soil age (biplot score = 0.55), bulk density (biplot score = −0.23), TC (biplot score = −0.15), and TN (biplot score = 0.07). On the vertical axis (CCA2, 20.03% of constrained variability), the most influential constraining variables were bulk density (biplot score = −0.82) and TC (biplot score = 0.81) followed by TN (biplot score = 0.71), soil age (biplot score = −0.51), and pH (biplot score = −0.28).

Association of bacterial diversity and composition with soil age and C accumulation rate. Using multiple alpha diversity indices, we assessed whether soil age and the C accumulation rate are significantly related to different aspects of microbial alpha diversity. However, we observed no significant relationships, positive or negative, among these parameters (all p > 0.05). Nonetheless, principal coordinate analysis

![Phylogenetic tree of the operational taxonomic units (OTUs) belonging to the top 10 genera.](image.png)

Figure 3. Phylogenetic tree of the operational taxonomic units (OTUs) belonging to the top 10 genera. Each color represents a particular genus. Each bar represents a particular OTU. The second levels indicate the relative abundance (logarithm) of the corresponding OTU. The highest relative abundance is 22.47%. Phylogenetic tree and bar chart were matched via Scalable Vector Graphics (1.1). Phylogenetic tree with bootstrap values can be found in Supplementary Figure 2. The relative abundance of each OTUs can be found in Supplementary Table 2.
(PCoA) of unweighted UniFrac distance ordinations did show that the bacterial community structures across the 33 soil samples were affected by soil age (Mantel test, $r = 0.195, p < 0.05$), but not the C accumulation rate ($r = -0.119, p = 0.889$) (Fig. 7). In regression analysis between the PCoA scores and soil ages, a significantly linear regression with the PCoA horizontal axis was found ($r = 0.701, p < 0.001$), which indicates that the variation in the bacterial community could be explained by soil age along the PCoA horizontal axis. This finding is consistent with the results of CCA analysis.

**Discussion**

We captured a rich microbial community from the peat cores collected from the Sanjiang Plain, at the southern edge of northern peatlands. However, we found that the bacterial community to be dominated by a few bacterial classes and OTUs that have frequently been detected in other peatland ecosystems. The most abundant phyla, Proteobacteria, Actinobacteria, Acidobacteria, Bacteroidetes and Firmicutes, have frequently been detected as dominant groups in other peatland ecosystems (oligo-mesotrophic, ombrotrophic and tundra)\(^5, 25, 26\).

Due to the vertical gradients in the soil properties we measured, such as TC, TN, pH, bulk density, and C/N ratio, as well as the availability of oxygen ($O_2$) and other electron acceptors, such as $SO_4^{2-}$ and $NO_3^-$, mentioned by others\(^5, 27\), the peat soil provides vertically stratified heterogeneous habitats for bacteria. In the present study, bacterial alpha diversity and composition varied as a function of peat depth and its associated physico-chemical properties, and pH, TC, TN, and bulk density showed significant effects on bacterial community structure.

The dominant microbial taxa detected at the surface are members of Acidobacteria, Gammaproteobacteria, Alphaproteobacteria (Acetobacteraceae), and Verrucomicrobia (Chthoniobacter). Acidobacteria, which are reportedly dominant among soils rich in organic matter, are involved in degradation of lignocellulose and cellulose\(^28-30\). A substantial number of carbohydrate-utilizing Acidobacteria and Alphaproteobacteria from acidic peatlands have been cultivated under aerobic conditions\(^31\). *Candidatus Solibacter*, the most abundant taxon among our samples, has been shown to be able to break down, utilize, and biosynthesize of diverse structural land-storage polysaccharides and exhibits resilience to fluctuating temperatures and nutrient-deficient conditions\(^32\).
Figure 5. A phylogenetic tree (a) and distribution patterns (b) of methanotrophs retrieved from fens S, H and Q.

Figure 6. Biplot of the CCA showing the relationships between community composition and soil properties. Arrows denote biplot scores for the constraining variables, with the longest independent variable vectors being most strongly related to variation in community composition (based on OTUs clustered at 97% identity). The circles represent individual sampling sites.
According to the research of Xing et al.17, potential peatlands cover an area of approximately 10,520 km² on the Sanjiang Plain and currently store ~0.26 Pg C. In response to future climate change and CO₂ fertilization, both net primary productivity and CH₄ fluxes will increase33. Methanotrophic bacteria can act as a natural barrier capable of significantly reducing the flux of methane into the atmosphere34, and bacteria that perform aerobic and anaerobic methanotrophy were detected in our samples. Although the relative abundance of type I and type II methanotrophs detected in the present study was similar to that in other peatlands25, 35, a broader sequence diversity (14 OTUs) of putative methanotrophs belonging to Alpha- and Gamma-proteobacteria was observed. The OTUs detected in this study, are most likely to act as key players in methane cycling in this environment. Further studies on the isolation and characterization of these microorganisms are needed and will facilitate our understanding of their physiological functions and ecological importance.

By comparing the AMS dating and soil properties of the cores of peat that developed at different ages, we found that soil properties, such as TN, TC, C accumulation rate and pH, were correlated with peat age. These peatlands initiated at different times, were affected by the climatic conditions, such as temperature and precipitation, etc, and underwent different initiation and decomposition processes, resulting in the respective soil properties. It is worth noting, that we also found that bacterial community turnover (beta diversity) to be significantly correlated with soil age, bacterial alpha diversity was not. We suggest that the effect of soil age on the bacterial community was due to its correlation with soil properties. The structure and activity of bacterial communities involved in C decomposition and release processes can directly affect the C accumulation in peatlands and other ecosystems. Nonetheless, when investigated the linkage of organic C accumulation and microbial community dynamics in a sandy loam soil, Zhang et al. found that C accumulation promoted the macroaggregation and reduced the effective diffusion coefficient of oxygen, causing changes in microhabitats and a shift in microbial communities36. We did not detect significant correlation between bacterial community turnover and the C accumulation rates. The role of bacteria in C cycling in these peatlands requires further verification via field experiments and laboratory ecophysiological studies.

In summary, we found that the peatlands on the Sanjiang Plain developed over a wide range of dates during the Holocene, with the peat cores being dated from 637 to 11,496 cal. yr BP. The relative abundance, distribution, and composition of the microbial communities in three different minerotrophic fens on the Sanjiang Plain, at the southern edge of northern peatlands, was investigated by next-generation sequencing. We captured a rich microbial community that included many rare OTUs but was dominated by a few bacterial taxa that have frequently been detected in other peatland ecosystems. Notably, a large diversity of methanotrophs affiliated with Alpha- and Gammaproteobacteria was detected. In addition, bacterial alpha diversity and composition varied as a function of peat depth and its associated physical-chemical properties, such as TC, TN, pH and bulk density. We also found that bacterial community turnover (beta diversity) to be significantly correlated with soil age, though this was not observed for bacterial alpha diversity.

Methods
Site description and sampling. We collected peat cores from three different minerotrophic fens, Shenjiadian (S), Honghe (H), and Qindelie (Q), on the Sanjiang Plain (129°11′–135°05′E, 43°49′–48°27′N), northeastern China (Fig. 1). The three fens began to develop during different periods of the Holocene26. This region is a large alluvial plain that is crossed by three major rivers, the Heilong River, Ussuri River and Songhua River, and has a total area of 10.9 × 10⁶ ha, an elevation of <200 m and a slope grade of <1:10,000. Over 70% of this region is dominated by freshwater wetlands that have developed in ancient riverbeds and waterlogged
depressions, and approximately 30% (or nearly 3.3 × 10^4 ha) of this region is covered by peatlands that developed under certain topographic conditions during the Holocene or earlier. The present climate of the plain is a temperate humid or subhumid continental monsoon climate. The mean annual temperature ranges from 1.4 to 4.3 °C, with an average maximum of 22 °C in July and an average minimum of −18 °C in January. The mean annual precipitation is 500–650 mm, and 80% of the rainfall occurs between May and September. In addition to precipitation, the fens receive water inputs from groundwater and they are primarily covered with sedges (Carex lasiocarpa). In May 2012, triplicate cores were collected from each fen using a Russian peat corer, the cores were subsampled for chronological, microbiological and physico-chemical analyses. The storage of the soil samples for chronological analysis is described in more detail in the work of Zheng et al. For the microbiological and physico-chemical analyses, peat cores were sectioned into depth intervals of 0 to 30, 30 to 60, 60 to 100 and 100 to 200 cm and; homogenized in sterile bags. After subsampling, the peat soils were stored in a sample incubator with a cooling function (2 to −10 °C) before being transferred to the laboratory. Chronological and physico-chemical characterization was conducted immediately after the samples arrived. Samples for DNA extraction were immediately frozen at −80 °C.

**Chronological and physico-chemical characterization.** Subsamples with a volume of 3 cm³ were used for loss on ignition (LOI) with sequential combustion at 500 °C to estimate the organic matter content. The bulk density of a 1-cm interval of each peat core was calculated according to the dry weight and volume of each subsample. Ash-free (organic matter) bulk density was calculated from the measurements of the bulk density and organic matter contents. All subsamples for AMS dating were dated with an accelerator mass spectrometry system at the Institute of Earth Environment, CAS. The AMS 14C dates were calibrated into calendar ages using the program Calib 7.02 based on the INTCAL 13 calibration dataset. Carbon accumulation rates were calculated by multiplying the organic carbon content (using 52% C in peat organic matter) by the bulk density and dividing by the age interval. Soil pH was measured in a 1:5 soil/water suspension. The total nitrogen (N) in the soil was determined by dichromate oxidation using a Continuous Flow Analytical System (SAN++, SKALAR, Netherlands).

**DNA extraction and sequencing.** Genomic DNA was extracted from 0.5 g of peat soil using a FASTDNA SPIN Kit for soil (MPBio, Santa Ana, USA) according to the manufacturer's instructions. The DNA concentrations were measured using a NanoDrop spectrophotometer (NanoDrop Technologies, Wilmington, USA) and diluted to 1 ng/µL. The V3-V4 region of the bacterial 16S rRNA gene was amplified using the primer pair 341F (CCTAYGGGRBGCASCAG), 806R (GGACTACNNGGGTATCTAAT) combined with Illumina adapter sequences, and barcodes. The specific barcodes used in this study are listed in Table 1. PCR reactions were performed in a 30-µL mixture containing 3 µL each primer (2 µM), 10 µL template DNA (1 ng/µL), 15 µL Phusion® High-Fidelity PCR Master Mix (BioLabs, Inc., New England, USA) and 2 µL water. The following thermal program was used for amplification: 95 for 1 min; followed by 30 cycles of 98 °C for 10 s, 50 °C for 30 s, and 72 °C for 30 s; and a final extension step at 72 °C for 5 min. Each sample was amplified in triplicate, and the PCR products were pooled and purified using QiaGen Gel Extraction Kit (Qiagen, Hilden, Germany). Amplicon-based sequencing libraries were generated using the TruSeq® DNA PCR Free Sample Preparation Kit (Illumina, San Diego, USA) according to the manufacturer’s instructions and pooled at an equimolar ratio. The Illumina HiSeq2000 platform at Novogene Bioinformatics Technology Ltd was used for 250-bp paired-end sequencing.

**Sequence data preprocessing and statistical analysis.** Raw sequences were divided into sample libraries via sample-specific barcodes and truncated after cutting off the barcode and the primer sequence. Forward and reverse reads with at least 10-bp overlaps and less than 5% mismatches were merged using FLASH. Quality filtering of the raw tags was performed according to the QIME (V1.7.0, http://qiime.org/index.html) quality control process, and all sequences shorter than 200 bp in length and with an average quality score lower than 25 in the raw reads were removed. The remaining sequences were subjected to chimera removal using the UCHIME algorithm (http://www.drive5.com/usearch/manual/uchime_algo.html). UPARSE (version 7.0.1001, http://drive5.com/uparse/) was employed to classify the operational taxonomic units (OTUs) at the 97% similarity level. The longest sequence that had the largest number of hits to other sequences in each OTU was screened as a representative sequence. All the OTUs with sequence numbers ≤ 2 were removed in the subsequent analysis. The RDP classifier (version 2.2, http://sourceforge.net/projects/rdp-classifier/) was used to annotate taxonomic information for each representative sequence. Representative sequences were aligned using the GreenGenes Database (version 2011, http://greengenes.lbl.gov/cgi-bin/nph-index.cgi), with a minimum identity of 80%. To study the phylogenetic relationships among different OTUs and the differences between the dominant species in different samples (groups), multiple sequence alignments were performed with the MUSCLE software (version 3.8.1, http://www.drive5.com/muscle/). For alpha and beta diversity analyses, the sequences were rarefied to depths of 21,811 sequences per sample, to minimize the effects of a different sampling efforts. Alpha diversity indices, including Chao1, Shannon, Simpson, ACE and PieLou equitability, were calculated. Beta diversity among the microbial communities was evaluated using both weighted and unweighted UniFrac distances. Evolutionary relationships were analyzed with neighbour-joining phylogenetic trees constructed in MEGA6. Pearson correlation analyses were applied to evaluate the relationships between the soil geochemical differences in soil properties across samples were determined by ANOVA in IBM SPSS (version 19.0, Chicago, IL, USA). Relationships between the taxonomic diversity and the soil properties were examined by linear regression analyses using SPSS. Principal coordinate analysis (PCoA) was conducted using the ape package (version 4.1) in R (version 3.3.3). Constrained correspondence analysis (CCA) was performed using the vegan package (version 2.4–1) in R. A Mantel test was used to test the relationship between bacterial similarities and overall environmental factors.
We also performed post hoc permutations using the function “envfit” in the vegan package to detect associations of the microbial community composition with environmental variables. A Mantel test was also used to examine the correlation between the UniFrac distances and soil age, and C accumulation rate. Heatmap plots were generated in R using the heatmap package51.

The Illumina sequencing data obtained in the present study have been deposited in the NCBI SRA database under the accession number SRP082472. The OTU table used for alpha and beta diversity analyses is provided as Supplementary Table 3.

References
1. Turunen, J., Tomppo, E. & Tolonen, K. Estimating carbon accumulation rate of undrained mires in Finland-application to boreal and subarctic regions. *Holocene* **12**, 69–80 (2002).
2. Yu, Z. Holocene carbon flux histories of the world's peatlands, global carbon cycle implications. *Holocene* **21**, 761–774 (2011).
3. Kurnianto, S. et al. Carbon accumulation of tropical peatlands over millennia: a modeling approach. *Global Change Biol.* **21**, 431–444 (2015).
4. Clymo, R. S., Turunen, J. & Tolonen, K. Carbon accumulation in peatland. *Oikos* **81**, 368–388 (1998).
5. Andersen, D., Chapman, S. J. & Artz, R. R. E. Microbial communities in natural and disturbed peatlands: A review. *Soil Biol. Biochem.* **57**, 979–994 (2013).
6. Bragazza, L., Parisi, J., Butler, A. & Bandgett, R. D. Biogeochemical plant-soil microbe feedback in response to climate warming in peatlands. *Nat. Clim. Chang* **3**, 273–277 (2013).
7. Kim, S. Y., Lee, S. H., Freeman, C., Fenners, N. & Kang, H. Comparative analysis of soil microbial communities and their responses to the short-term drought in bog, fen, and riparian wetlands. *Soil Biol. Biochem.* **40**, 2874–2880 (2008).
8. Potila, H. & Sarjala, T. Seasonal fluctuation in microbial biomass and activity along a natural nitrogen gradient in a drained peatland. *Soil Biol. Biochem.* **51**, 20–34 (2012).
9. Lin, X. I., Nielson, K., Kennedy, D., Fredrickson, J., Bjornstad, B. & Konopka, A. Vertical stratification of subsurface microbial community composition across geological formations at the Hanford Site. *Appl. Environ. Microbiol.* **14**, 414–425 (2012).
10. Bae, H. & Hou, A. 23S rRNA gene-based enterococci community signatures in Lake Pontchartrain, Louisiana, USA, following urban runoff inputs after hurricane Katrina. *Microb. Ecol.* **65**, 289–301 (2013).
11. Lin, X. et al. Microbial community structure and activity linked to contrasting biogeochemical gradients in bog and fen environments of the Glacial Lake Agassiz Peatland. *Appl. Environ. Microbiol.* **7033–7031 (2012)."
12. Morris, S. A., Radajewski, S., Willison, T. W. & Murrell, J. C. Identification of the functionally active methanotroph population in a peat soil microcosm by stable isotope probing. *Appl. Environ. Microbiol.* **68**, 1446–1453 (2002).
13. Jaatinen, K., Tuittila, E. S., Laine, J., Yrjälä, K. & Fritze, H. Methane-oxidising bacteria in a Finnish raised mire complex: effects of site fertility and drainage. *Microb. Ecol.** **50**, 429–439 (2005).
14. Huang, J. Y. & Song, C. Effects of land use on soil water soluble organic C and microbial biomass C concentrations in the Sanjiang Plain in northeast China. *ACTA Arg. Scand. B-S P** **60**, 182–188 (2010).
15. White, M. L., Varner, R. K., Crill, P. M. & Mosedale, C. H. Control on the seasonal exchange of CH3Br in temperate peatlands. *Soil Biol. Biochem.* **85**, 2046–2056 (2015).
16. Sun, L. C., Brauer, L. S., Cadillo-Quiroz, H., Zinder, H. S. & Yavitt, B. J. Seasonal Changes in Methanogenesis and Methanogenic Community in Three Peatlands, New York State. *Front Microbiol* **3**, 81 (2012).
17. Lin, X. et al. Holocene peatland initiation and carbon storage in temperate peatlands of the Sanjiang Plain, Northeast China. *Soil Biol. Biochem* **26**, 70–79 (2016).
18. Yu, Z., Beilman, D. W. & Jones, M. C. Sensitivity of northern peatland carbon dynamics to Holocene climate change in Carbon Cycling in Northern Peatlands: Geophysical Monograph. (ed. Baird A. J. et al.) 55–69 (American Geophysical Union, 2009).
19. Lin, X. I., Liu, X. T. & Wang, R. F. China’s wetlands and agroecological engineering. *Ecol. Eng.* **2**, 291–301 (1993).
20. Zhang, Z. et al. The peatlands developing history in the Sanjiang Plain, NE China, and its response to East Asian monsoon variation. *Sci. Rep.* **5**, 11316 (2015).
21. Shang, Q. et al. Illumina-based analysis of the rhizosphere microbial communities associated with healthy and wilted Lanzhou lily (*Lilium davidii var. unicolor*) plants grown in the field. *World J. Microbiol. Biotechnol.* **32**, 95 (2016).
22. Potila, H. & Sarjala, T. Seasonal fluctuation in microbial biomass and activity along a natural nitrogen gradient in a drained peatland. *Soil Biol. Biochem.* **36**, 1047–1055 (2004).
23. O’Brien, L. S. et al. Spatial scale drives patterns in soil bacterial diversity. *Environ. Microbiol.* **18**, 2039–2051 (2016).
24. Thomann, M. N., Currah, R. S. & Bayley, S. E. Patterns of distribution of microfungi in decomposing bog and fen plants. *Can. J. Bot.* **82**, 710–720 (2004).
25. Lin, X. I. et al. Microbial Community Stratification Linked to Utilization of Carbohydrates and Phosphorus Limitation in a Boreal Peatland at Marcell Experimental Forest, Minnesota, USA. *Appl. Environ. Microbiol.* **80**, 3518–3530 (2014).
26. Pankratov, T. A., Belova, S. E. & Dedvysh, S. N. Evaluation of the phylogenetic diversity of prokaryotic microorganisms in Sphagnum peat bogs by means of fluorescence in situ hybridization (FISH). *Microbiology* **74**, 722–728 (2006).
27. Deng, J. et al. Shifts of tundra bacterial and archaeal communities along a permafrost thaw gradient in Alaska. *Mol. Ecol.** **24**, 222–234 (2015).
28. Eichorst, S. A., Kuske, C. R. & Schmidt, T. M. Influence of plant polymers on the distribution and cultivation of bacteria in the phylum Acidobacteria. *Appl. Environ. Microbiol.* **77**, 586–596 (2011).
29. Pankratov, T. A., Ivanova, A. O., Dedvysh, S. N. & Liesack, W. Bacterial populations and environmental factors controlling cellulose degradation in an acidic Sphagnum peat. *Environ. Microbiol.* **13**, 1800–1814 (2011).
30. Ward, N. L. et al. Three genomes from the phylum Acidobacteria provide insight into the lifestyles of these microorganisms in soils. *Appl. Environ. Microbiol.* **75**, 2046–2056 (2009).
31. Dedvysh, S. N. Cultivating uncultured bacteria from northern wetlands: knowledge gained and remaining gaps. *Front Microbiol.* **2**, 184 (2011).
32. Rawat, R. S., Mannisto, K. M., Bromberg, Y. & Haggblom, M. M. Comparative genomic and physiological analysis provides insights into the role of Acidobacteria in organic carbon utilization in Arctic tundra soils. *FEBS Microbiol. Ecol.* **82**, 341–353 (2012).
33. Li, T. et al. Prediction CH4 Emissions from the Wetlands in the Sanjiang Plain of Northeastern China in the 21st Century. *PLOS ONE* **11**, e0158872 (2016).
34. Danilova, O. V., Belova, S. E., Gagarinova, I. V. & Dedvysh, S. N. Microbial Community Composition and Methanotrophic Diversity of a Subarctic Wetland in Russia. *Microbiology* **85**, 583–591 (2016).
35. Kip, N. et al. Global prevalence of methane oxidation by symbiotic bacteria in peat- moss ecosystems. *Nat. Geosci.* **3**, 617–621 (2010).
36. Zhang, H., Ding, W., Yu, H. & He, X. Linking organic carbon accumulation to microbial community dynamics in a sandy loam soil: result of 20 years compost and inorganic fertilizer repeated application experiment. *Biol Fertil Soils* **51**, 137–150 (2015).
37. Liu, X. T. Wetland and its rational utilization and conservation in the Sanjiang Plain (in Chinese) (Jilin Science Technology Press,1995).
38. Ma, X. H., Liu, X. T. & Wang, R. F. China’s wetlands and agro-ecological engineering. *Ecol. Eng.* **2**, 291–301 (1993).
39. Song, C. C., Xu, X. F., Tian, H. Q. & Wang, Y. Y. Ecosystem-atmosphere exchange of CH₄ and N₂O and ecosystem respiration in wetlands in the Sanjiang Plain, Northeastern China. *Global Change Biol.* **15**, 692–705 (2009).
40. Dean, W. E. Determination of carbonate and organic matter in calcareous sediments and sedimentary rocks by loss on ignition comparison with other methods. *J. Sediment. Res.* **44**, 242–248 (1974).
41. Stuiver, M. & Reimer, P. J. Extended ¹³C data base and revised CALIB 3.0 ¹³C age calibration program. *Radiocarbon* **38**, 215–230 (2006).
42. Vitt, D. H., Halsey, L. A., Bauer, I. E. & Campbell, C. Spatial and temporal trends in carbon storage of peatlands of continental western Canada through the Holocene. *Can. J. Earth Sci.* **37**, 683–693 (2000).
43. Zhang, B. *et al.* Responses of soil microbial communities to experimental warming in alpine grasslands on the Qinghai-Tibet Plateau. *PLOS ONE* **9**, e103859 (2015).
44. Caporaso, J. G. *et al.* Ultra-high-throughput microbial community analysis on the Illumina HiSeq and MiSeq platforms. *ISME J.* **6**, 1621–1624 (2012).
45. Bokulich, A. N. *et al.* Quality-filtering vastly improves diversity estimates from Illumina amplicon sequencing. *Nat. methods* **10**, 57–59 (2013).
46. Edgar, R. C. MUSCLE: multiple sequence alignment with high accuracy and high throughput. *Nucleic Acids. Res.* **32**, 1792–1797 (2004).
47. DeSantis, T. Z. *et al.* Greengenes, a chimera-checked 16S rRNA gene database and workbench compatible with ARB. *Appl Environ. Microbiol.* **72**, 5069–5072 (2006).
48. Oksanen, J. *et al.* The vegan package. *Community ecology package: 631–637* (2007).
49. Banerjee, S. *et al.* Network analysis reveals functional redundancy and keystone taxa amongst bacterial and fungal communities during organic matter decomposition in an arable soil. *Soil Biol. Biochem.* **97**, 188–198 (2016).
50. Paradis, E., Claude, J. & Strimmer, K. APE: analyses of phylogenetics and evolution in R language. *Bioinformatics* **20**, 289–290 (2004).
51. Zhang, N. *et al.* Precipitation modifies the effects of warming and nitrogen addition on soil microbial communities in northern Chinese grasslands. *Soil Biol. Biochem.* **89**, 12–23 (2015).

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**Author Contributions**

Xue Zhou conducted all the experiments except for the chronological characterization of the soil, and wrote the paper with the help of Zhenqing Zhang, Xiujun Li and Chunjie Tian. Zhenqing Zhang collected the peat cores and designed and performed the chronological characterization of the soil. Lei Tian carried out some of the data analysis. Chunjie Tian helped in the research design and, revised the paper and is responsible for the funds supporting this study. All authors reviewed the manuscript.

**Additional Information**

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