Different Responses of Soil Bacterial Communities to Nitrogen Addition in Moss Crust

Tingwen Huang1, Weiguo Liu1*, Xi-En Long2, Yangyang Jia1, Xiyuan Wang1 and Yinguang Chen3*

1 Ministry of Education, Key Laboratory of Oasis Ecology, College of Resources and Environment Science, Xinjiang University, Urumqi, China, 2 School of Geographic Sciences, Nantong University, Nantong, China, 3 College of Environmental Science and Engineering, Tongji University, Shanghai, China

Bacterial communities in soil serve an important role in controlling terrestrial biogeochemical cycles and ecosystem processes. Increased nitrogen (N) deposition in Northwest China is generating quantifiable changes in many elements of the desert environment, but the impacts of N deposition, as well as seasonal variations, on soil bacterial community composition and structure are poorly understood. We used high-throughput sequencing of bacterial 16S rRNA genes from Gurbantüngü Desert moss crust soils to study the impacts of N addition on soil bacterial communities in March, May, and November. In November, we discovered that the OTU richness and diversity of soil bacterial community dropped linearly with increasing N input. In November and March, the diversity of the soil bacterial community decreased significantly in the medium-N and high-N treatments. In May, N addition caused a substantial change in the makeup of the soil bacterial composition, while the impacts were far less apparent in November and March. Furthermore, the relative abundance of major bacterial phyla reacted non-linearly to N addition, with high-N additions decreasing the relative richness of Proteobacteria, Bacteroidetes, and Acidobacteria while increasing the relative abundance of Actinobacteria and Chloroflexi. We also discovered that seasonality, as characterized by changes in soil moisture, pH, SOC, and AK content, had a significant impact on soil bacterial communities. Significant variations in the makeup of the community were discovered at the phylum and genus levels throughout the various months. In May, the variety of soil bacterial community was at its peak. Further investigation showed that the decrease in soil bacterial diversity was mostly attributed to a drop in soil pH. These results indicated that the impact of N deposition on the soil bacterial community was seasonally dependent, suggesting that future research should evaluate more than one sample season at the same time.

Keywords: nitrogen deposition, moss crusts, seasonal changes, bacterial composition, community diversity

INTRODUCTION

Biological soil crusts (BSCs) are common in arid ecosystems. They are made up of mosses, lichens, cyanobacteria, algae, fungus, and bacteria and help to keep the soil wet, prevent erosion, and promote plant development (Belnap and Lange, 2001). Bacteria make up the majority of BSCs and play an important role in their development and persistence (Wang et al., 2015). Meanwhile, bacteria have been recognized for their important role in the breakdown of organic materials in the
desert environment (Moquin et al., 2012). The quantity, variety, and function of the soil bacterial community, on the other hand, may change in response to external environmental variables such as N deposition (Wang et al., 2018), the dry-wet cycle (Angel and Conrad, 2013), and temperature (Abed et al., 2015). Thus, studying changes in the soil bacterial community may aid in better understanding of the impacts of environmental change on soil ecosystem processes.

Anthropogenic N deposition (mostly nitrogen oxide and ammonia) has more than tripled in the last five decades as a result of agricultural fertilizer and fossil fuel burning (Bobbink et al., 2010; Galloway et al., 2014). The rate of N deposition is expected to rise steadily and may double by 2050 (Phoenix et al., 2012). Empirical research has revealed that N deposition has a significant impact on the bacterial community through plant-soil feedbacks (Li H. et al., 2016; Zeng et al., 2016). For example, N deposition may increase plant production, reduce root development, change bacterial respiratory substrates, and therefore influence bacterial responses to N deposition (LeBauer and Treseder, 2008; Treseder, 2008). Moreover, many studies have shown that the effects of N deposition on soil bacteria are affected by soil characteristics (Jia et al., 2021; Yue et al., 2021).

Previous meta-analysis showed that N deposition reduced soil bacterial diversity and the relative abundance of Actinobacteria and Nitrospirae via lowering soil pH and changing soil nutritional conditions (Wang et al., 2018). Similarly, one recent study found that N addition-induced acidity was the primary driver to the decline in soil bacterial diversity (Yang et al., 2020). Nonetheless, Liu et al. (2020) discovered that the response of bacterial community composition and diversity to N deposition had a concentration threshold, beyond which greater N concentration inhibited bacterial diversity. DeForest et al. (2004), on the other hand, found that increased N deposition had no impact on the abundance or composition of bacteria in a hardwood forest. These inconsistencies suggested that the reactions of the soil bacterial community to N deposition are still unknown and need more investigation.

Understanding the dynamics of the desert soil bacterial population is critical for long-term desert management and accurate forecasting of desert ecosystem conditions in the face of future environmental change scenarios (Cregger et al., 2012; Jia et al., 2021). Soil bacterial communities are known to be influenced by seasonal changes in soil moisture and temperature (Liu et al., 2018; Wang et al., 2020). The organization of bacterial communities in dryland ecosystems were shown to be strongly related to soil moisture content (Angel and Conrad, 2013; Wu et al., 2020; Jia et al., 2021). Furthermore, the soil bacterial community has undergone substantial seasonal variations as a result of temperature (Abed et al., 2015). During warm times, the soil bacterial community richness index (ACE and Chao1) and diversity index (Shannon) values were the greatest (i.e., summer; Jia et al., 2020).

Moss crust, which is common in the Gurbantünggüt Desert, is considered the most mature crust type (Belnap and Lange, 2001). Moss crusts, on the other hand, are very susceptible to environmental changes owing to their basic structure and absence of a cuticular barrier (Juha et al., 2015). In recent decades, the Gurbantünggüt Desert has seen increased N deposition, as well as significant seasonal fluctuations in soil temperature and moisture (Li et al., 2014). Although the effects of N deposition on soil bacterial communities dominated by lichens or cyanobacteria have been studied fairly thoroughly (Mueller et al., 2015; Wang et al., 2015), the effects of N deposition in combination with seasonal changes on soil bacterial communities in moss crusts remain unknown. Thus, we examined the soil bacterial community in moss crust, Gurbantünggüt Desert, in a controlled experiment plot that received N addition at four different levels (0, 1.8, 3.6, and 7.2 g N m⁻² year⁻¹). The effects of various amounts of N on the community composition and diversity of moss crust soil bacteria throughout seasons were investigated in this research. We predicted that (1) N addition would reduce soil bacterial community diversity and alter bacterial community composition, and (2) N addition’s impacts would be seasonal, with less influence in May when circumstances are hot and dry.

MATERIALS AND METHODS

Sample Sites
A simulated N deposition experiment was carried out in the southern edge of the Gurbantünggüt Desert (44°11′–46°20’N, 84°31′–91°20’E), China’s biggest permanent and semi-fixed desert (4.88 × 10⁴ km²), situated in the Jungger Basin, Xinjiang, at a height of 400–600 m. The climate in the area is characteristic of a continental dry environment, with an annual average temperature of 6.6°C and a maximum temperature of more than 40°C. The annual mean precipitation is 70–150 mm, with the most of it falling in the spring, while the yearly evaporation is 3,000–3,500 mm. In the winter, there is a thick snow cover of more than 20 cm, and the snow cover is constant for approximately 95–110 days (Hui et al., 2016). The desert contains a variety of semi-fixed sand dunes ranging in height from 10 to 50 meters and lengths ranging from tens to hundreds of kilometers. The primary vegetation consists of four shrub species, Haloxylon ammodendron, Haloxylon persicum, Ephedra distachya, and Ceratoides latens, interspersed with patches of well-developed BSCs, and the moss crusts are mostly found on the inter-dunes, plant interspace, and underplant canopies.

Experimental Design
Twelve experimental plots were constructed across a 100 × 100 m area based on the distribution of moss crusts over interdunes. Each plot (5 × 5 m) was randomly assigned one of four N levels. Each plot was surrounded by a 25 cm stainless steel frame, 20 cm of which was buried in the soil. External environmental variables were eliminated by removing any interfering plants from the plots, and the dead branches and debris covering the plots and adjacent regions were cleaned on a regular basis. Considering the background level (3.6 g N m⁻² year⁻¹) of atmospheric N deposition in Xinjiang over the last 10 years (Huang et al., 2015), N addition was applied at four levels in our experiment: control (N0, 0 g N m⁻² year⁻¹), low-N (LN, 1.8 g N m⁻² year⁻¹), medium-N (MN, 3.6 g N m⁻² year⁻¹), and high-N (HN, 7.2 g N m⁻² year⁻¹), and each N treatment had three replicate plots.
the end of March 2017, NH$_4$NO$_3$ was diluted in 1 L of deionized water (equal to 1 mm precipitation) and uniformly sprayed over the LN, MN, and HN plots in each block. Each control (N0) plot received 1 L of deionized water. We performed the N addition in March because agricultural activity in the adjacent study area peaks in March, when a significant quantity of N fertilizer is utilized, potentially increasing N emission and deposition on desert lands (Zhou et al., 2012). Furthermore, by March, the snow in the study region had almost entirely melted.

**Soil Sample Collection**

We collected soil from the research region on the 25th of May 2017 (May, the short-lived plant growth season), the 10th of November 2017 (November, the soil freezing time), and the 16th of March 2018 (March, the snowmelt period). Before taking the cores, the moss crust layer was removed from the soil surface. Using a sterile trowel, five soil cores (with a depth of 5 cm and a diameter of 3.5 cm) were collected from five random locations on each plot. To reduce geographic variability, the five soil cores were combined into a single sample. Finally, we collected 36 soil samples (4 N levels × 3 repetitions × 3 months). Soil samples were placed into a portable chilled box and transferred to the laboratory after plant roots and big stones were removed using a 2 mm screen. Each sample was then subdivided into two sub-samples, one of which was kept at −80°C and used to extract soil DNA and gain information on bacterial community structure, and the other of which was air-dried to assess soil physicochemical characteristics.

**Physicochemical Analysis of Soil Samples**

Soil organic carbon (SOC) was estimated by using the potassium dichromate-sulfuric acid oxidation method; total nitrogen (TN) was determined following Kjeldahl digestion on a Nitrogen Analyzer System (KJELTEC 2300 AUTO SYSTEM II, Foss Tecator AB, Höganäs, Sweden). Available nitrogen (AN) was estimated by the alkali-hydrolysis digestion method. Total phosphorus (TP) was determined by the H$_2$SO$_4$-HClO$_4$ digestion method. Available phosphorus (AP) data was determined by the NaHCO$_3$ extraction molybdenum-antimony colorimetric method. Soil samples were brought to the laboratory and mixed with water at a ratio of 1:2.5 to measure the pH using a compound electrode. Soil NO$_3$-N content was extracted with 2 M KCl using a continuous flow analytical system (Santt System; Skalar, MA, United States) and Agilent Bio analyzer 2100 system (Agilent Technologies, Santa Clara, CA, United States), and then the library was sequenced on an Illumina HiSeq platform (Novogene, Beijing, China). The raw FASTQ files have been deposited in the SRA of the NCBI database under the SRA accession: PRJNA701371.

Raw FASTQ files were demultiplexed by QIIME (QIIME 1.9.1), quality-filtered by Trimmomatic, and merged by FLASH (v1.2.7, Magoc and Salzberg, 2011) with the following criteria: (i) the 250 bp reads were truncated at any site receiving an average quality score of <20 over a 10 bp sliding window, and the truncated reads <50 bp were removed. (ii) exact barcode matching, reads containing ambiguous characters, and more than two base mismatch in primer matching were discarded. (iii) only overlapping sequences >10 bp were assembled according to their overlapped sequence, the reads that could not be assembled were discarded (Caporaso et al., 2010). (iv) the sequences were assigned to each sample according to the primers and 6-bp barcode. Then, the barcodes were trimmed and the sequence direction was flipped.

Operational taxonomic units (OTUs) were clustered at a 97% similarity level cutoff using UPARSE (version 7.1), and chimeric sequences were identified and removed using UCHIME (Edgar et al., 2011). The representative sequences were selected for each OTU and used the Ribosomal Database Project (RDP) classifier$^3$ to annotate taxonomic information for each representative sequence, and the confidence threshold was set to 80% (Wang et al., 2007). The sampling depth for rarefaction curve analysis was optimized to 13,025 sequences for soil bacterial. Alpha biodiversity was evaluated using the abundance-based indices of OTU richness, Pielou index, and

1. [http://ccb.jhu.edu/software/FLASH/](http://ccb.jhu.edu/software/FLASH/)
2. [http://drive5.com/uparse/](http://drive5.com/uparse/)
3. [http://rdp.cme.Msu.edu/classifier/classifier.jsp](http://rdp.cme.Msu.edu/classifier/classifier.jsp)

**Library Preparation and Sequencing**

Sequencing libraries were implemented using the Illumina TruSeq DNA PCR-Free Library Prep Kit (Illumina, United States) according to the manufacturer's protocols and index codes were added. Library quality was quantified and monitored on the Qubit® 2.0 Fluorometer (Thermo Scientific, Waltham, MA, United States) and Agilent Bio analyzer 2100 system (Agilent Technologies, Santa Clara, CA, United States), and then the library was sequenced on an Illumina HiSeq platform (Novogene, Beijing, China). The raw FASTQ files have been deposited in the SRA of the NCBI database under the SRA accession: PRJNA701371.

**DNA Extraction From Soil Samples and Quantitative PCR Analysis**

Total genomic DNA was extracted from 0.5 g soil samples using the TIANamp Soil DNA Kit (TIANGEN BioTech (Beijing) Co., Ltd) following the manufacturer's instructions. The concentration and purity of DNA were monitored on 1% agarose gels. The DNA concentration was adjusted to 1 ng/µL using sterile water. The V3-V4 region of the bacterial 16S rRNA gene was amplified by PCR (95°C for 1 min, followed by 30 cycles of 95°C for 10 s, 50°C for 30 s, 72°C for 30 s, and a final extension cycle at 72°C for 5 min), using the 341F primers 5’-GCTAYGGRBGCASCAG-3’ and 806R 5’-GGACTACNNGGATATCTAAAT-3’ (Xia et al., 2015), each sample former primer contains a unique 6 bp Barcode (Bensky et al., 1993). PCR reactions were performed in a 50 µL mixture containing 10 µL of the target DNA extract, 15 µL of Phusion Master Mix (2×, New England Biolabs), 2 µL of sterile water, 3 µL of primer, 0.2 µM of forward and reverse primers, and about 10 ng template DNA. PCR products were detected by 2% agarose gel electrophoresis and target DNA bands were purified using a GeneJET Gel Extraction Kit (Thermo Scientific).
Shannon index, calculated in QIIME2, to analyze the complexity of bacterial diversity in each sample.

**Data Analysis**

One-way analysis of variance (ANOVA) was used to analyze the effect of N addition on soil parameters and bacterial community. Two-way ANOVA was used to determine the effects of N addition and seasonal change on soil environmental parameters and bacterial community. Heatmap of the dominant phyla (or genera) and soil physicochemical properties in each sample were drawn using the R packages (pheatmap, psych, reshape2). Principal coordinate analysis (PCoA), based on Bray-Curtis distances of the OTU level data, was performed to interpret the relative similarity of the bacterial community from each sample. Redundancy analysis (RDA; post hoc permutation tests with 999 permutations) was used as the statistical criterion to assess multivariate changes in bacterial community composition among treatments using the “vegan” package in R version 4. 0. 2, the most discriminating environmental factors were selected by the “forward selection” of ordiR2step. Additionally, structure equation models (SEMs) were used to explain how N addition and the seasonal change influences soil bacterial community composition/diversity through hypothetical pathways, the response variables for diversity and composition are Shannon diversity index and community OTUs, respectively. The response variables for soil bacterial community were indicated by principal component scores (both PC1 scores were used) (Wu et al., 2019), which are usually used in SEM analyses. The SEMs were analyzed using IBM SPSS Amos 24, and Adobe Illustrator CC 2019 was used to process the graphics.

**RESULTS**

**Effects of N Addition and Seasonal Changes on Soil Properties**

Soil AP, AN, TP, and TN content showed increasing trends with N addition, but pH showed a decreasing trend with N addition (Figure 1). Soil NO$_3^-$-N concentrations and AK content both showed significant variations between N concentrations ($P < 0.05$). In May, SOC content dropped linearly across the gradient of N inputs. Simultaneously, we discovered substantial changes in soil pH, moisture, and SOC concentration between sample months (Figure 1; Table 1). In May, SOC content was lowest, whereas soil pH was highest ($P < 0.05$). Soil moisture was highest in November (21%), followed by March (13%) and May (11%, $P < 0.05$). Furthermore, we also found significant interactions between N addition and seasonal changes in soil pH, TP, AN, SOC, NO$_3^-$-N, and AK (Table 1).

**Responses of Soil Bacterial Community Diversity to N Addition and Seasonal Change**

In May, no significant changes in OTU richness, Shannon, or Pielou index were found as N increased (Figure 2). In November, however, the OTU richness and diversity of bacterial community dropped linearly with increasing N inputs. Simultaneously, MN and HN addition showed a substantial detrimental impact on community diversity in both November and March ($P < 0.05$). Furthermore, the bacterial diversity indices varied between sample months. In May, the soil bacterial community was at its most diverse, with an OTU richness index of 6,029–8,920, a Pielou index of 0.77–0.78, and a Shannon index of 9.70–10.28 (Figure 2; Supplementary Table 1). However, no significant impact was observed in the interaction N treatment seasonal change for bacterial community diversity (Table 2).

**Effects of Increased N Addition on Soil Bacterial Community Structure in Different Seasons**

Proteobacteria, Actinobacteria, Bacteroidetes, Acidobacteria, and Chloroflexi were the most prevalent phyla across all N deposition levels and sample months, accounting for 78–93% of total bacterial sequences (Figure 3). Some bacterial phyla were shown to be more susceptible to N addition in May than in November and March (Figure 3). For example, in N-amended soils, the relative abundances of Actinobacteria decreased by 6.7 and 6.8% for LN and MN, respectively, but rose by 7.1% for HN. The effects of N addition on the relative abundance of Acidobacteria, on the other hand, exhibited the opposite trend. The relative abundance of Chloroflexi rose substantially after HN treatment (15.84%). When compared to the N0 treatment, the relative abundance of Proteobacteria was higher in the LN (37.97%) and lower in the MN (31.38%) and HN (28.01%) treatments. Furthermore, the five most prevalent genus (>1%) across all N addition treatments were identified (Supplementary Figure 1). The variations in dominating genera were also seen across N treatments. Under N addition, the relative abundances of Sphingomonas and Balneimonas changed substantially in May and November, whereas Kaistobacter changed significantly in all three sample months.

Soil bacteria’s relative abundance changed seasonally. At the phyllum level, Proteobacteria, Bacteroidetes, and TM7 had the highest relative abundance in November compared to March and May, while Actinobacteria, Acidobacteria, and Chloroflexi had the lowest. Firmicutes, Gemmatimonadetes, Planctomycetes, and Verrucomicrobia were more abundant in March than in the previous 2 months. In November, the genera Sphingomonas, Kaistobacter, Massilia, and Adhaeribacter had the greatest relative abundance. Rubrobacter’s relative abundance was in the sequence of May > March > November. Furthermore, we discovered that N addition and seasonal variation had an interaction impact on the relative abundance of several prominent phyla and genera, including Actinobacteria, Acidobacteria, Chloroflexi, Planctomycetes, Sphingomonas, Balneimonas, Kaistobacter, and Geodermatophilus (Table 2).

Bray-Curtis distances were calculated using Principal coordinate analysis (PCoA) to show the differences in bacterial communities across treatments (Figure 4). PCoAs 1 and 2 explained 41.53 and 19.92% of the total variance, respectively. The bacterial communities were divided based on the months rather than the amounts of N addition. This meant that the
Huang et al. Nitrogen Alter Soil Bacterial Community

FIGURE 1 | Seasonal changes in soil physicochemical properties in moss crust soil with different levels of N additions. N0 = 0 g N m\(^{-2}\) year\(^{-1}\); LN = 1.8 g N m\(^{-2}\) year\(^{-1}\); MN = 3.6 g N m\(^{-2}\) year\(^{-1}\); HN = 7.2 g N m\(^{-2}\) year\(^{-1}\). Different capital letters indicate a significant difference among the three sampling months with the same concentration; different lowercase letters indicate a significant difference among four N treatments in the same sampling month. Vertical bars show the standard error (SE) \((n = 3)\).

TABLE 1 | Effects of nitrogen addition, seasons, and their interaction on soil variables, as indicated by two-way ANOVA statistics.

| Factors | Nitrogen | Season | Nitrogen x Season |
|---------|----------|--------|------------------|
|         | F        | P      | F                | P      |
| pH      | 0.482    | 0.037  | 18.328           | < 0.001|
| Soil moisture | 0.010    | 0.999  | 972.799          | < 0.001|
| TP (mg/g) | 13.184   | 0.002  | 0.210            | 0.815  |
| TN (mg/g) | 36.194   | < 0.001| 0.235            | 0.795  |
| AN (mg/kg) | 30.356   | < 0.001| 0.260            | 0.776  |
| AP (mg/kg) | 14.094   | 0.001  | 0.842            | 0.462  |
| SOC (g/kg) | 0.153    | 0.925  | 22.497           | < 0.001|
| NO\(^3\)-N (g/kg) | 5.669    | 0.022  | 0.180            | 0.838  |
| AK (g/kg) | 2.259    | 0.159  | 1.289            | 0.322  |

Values represent mean with standard error in parenthesis. TP, total phosphorus; TN, total nitrogen; AN, available nitrogen; AP, available phosphorus; SOC, soil organic carbon; NO\(^3\)-N, nitrate nitrogen; AK, available potassium.

Relationships Between the Bacterial Community Structure and Soil Properties

The composition of the soil bacterial community was significantly different in different months based on the RDA analysis, which Axis 1 explained 55.79% of the total variance, and Axis 2 explained 14.74% (Figure 5). The result revealed that the soil bacterial community was significantly related to soil moisture \((F = 49.453, P < 0.001)\), SOC \((F = 15.17, P < 0.001)\), pH \((F = 5.162, P = 0.007)\), and AK \((F = 4.169, P = 0.012)\). Correlation analysis revealed that the bacterial phylum Actinobacteria and Chloroflexi were positively correlated with soil pH, and negatively correlated with soil moisture, SOC, and AK contents (Figure 6A). Conversely, Gemmatimonadetes, Firmicutes, Planctomycetes, and Verrucomicrobia were negatively correlated with soil pH impacts of season on soil bacterial composition were greater than the effects of N addition.
Huang et al. Nitrogen Alter Soil Bacterial Community

FIGURE 2 | Seasonal changes in soil bacterial OTU richness and diversities in moss crusts with different levels of N additions. OTU richness (A), Pielou index (B), and Shannon index (C). N0 = 0 g N m⁻² year⁻¹; LN = 1.8 g N m⁻² year⁻¹; MN = 3.6 g N m⁻² year⁻¹; HN = 7.2 g N m⁻² year⁻¹. Different capital letters indicate a significant difference among the three sampling months with the same concentration of N; different lowercase letters indicate a significant difference among four N treatments in the same sampling month. Vertical bars show the standard error (SE) (n = 3).

TABLE 2 | Effects of nitrogen addition, season, and their interactions on soil bacterial diversity as indicated by two-way ANOVA statistics, based on the relative abundance of phyla level.

| Factors          | Nitrogen  | Season | Nitrogen × Season |
|------------------|-----------|--------|-------------------|
|                  | F   | P      | F   | P      | F   | P      |
| OTU richness     | 0.449 | 0.725  | 11.634 | 0.002  | 0.694 | 0.627  |
| Shannon index    | 0.893 | 0.485  | 78.595 | <0.001 | 1.230 | 0.342  |
| Pielou index     | 1.701 | 0.243  | 170.318 | <0.001 | 1.183 | 0.365  |
| Proteobacteria   | 0.053 | 0.983  | 177.079 | <0.001 | 2.384 | 0.077  |
| Actinobacteria   | 1.784 | 0.228  | 124.896 | <0.001 | 6.631 | 0.001  |
| Bacteroidetes    | 0.828 | 0.514  | 29.450 | <0.001 | 1.018 | 0.450  |
| Acidobacteria    | 3.253 | 0.081  | 20.404 | <0.001 | 4.298 | 0.009  |
| Chloroflexi      | 3.622 | 0.065  | 134.703 | <0.001 | 7.034 | 0.043  |
| Firmicutes       | 0.866 | 0.497  | 45.454 | <0.001 | 1.049 | 0.431  |
| Cyanobacteria    | 2.782 | 0.110  | 10.267 | 0.008  | 3.250 | 0.081  |
| Gemmatimonadetes | 2.956 | 0.098  | 70.792 | <0.001 | 1.250 | 0.333  |
| Planctomycetes   | 1.227 | 0.362  | 33.762 | <0.001 | 4.428 | 0.041  |
| TM7              | 1.807 | 0.224  | 124.052 | <0.001 | 3.319 | 0.078  |
| Verrucomicrobia  | 4.328 | 0.043  | 9.887  | 0.002  | 1.242 | 0.318  |

and soil moisture, but positively correlated with soil SOC and AK contents. There was a positive correlation between Proteobacteria, Bacteroidetes, and TM7 and soil moisture, SOC, and AK contents. At the genus level, we found that Afifella exhibited significant positive correlations to soil pH. Rubrobacter exhibited highly negative correlations with soil moisture and SOC content, while Sphingomonas, Adhaeribacter, Kaistobacter, Rhodocytophaga, Massilia, and Ramlibacter correlated positively with soil moisture and SOC content (Figure 6B). Structure equation models analysis also showed that N addition significantly affected soil N and pH, and explained 96 and 26% of the total variance, respectively (Figure 7A). Soil pH explained the total variance in the soil bacterial community diversity (56%) and composition (75%) (Figure 7A), while soil N had no directly affected the soil bacterial community, these results shown that increased N mainly affects soil bacterial communities by causing soil acidification. Seasonal changes significantly affected soil moisture, SOC, and pH, and
explained 55, 82, and 76% of the total variance (Figure 7B), respectively. Three soil parameters explained 67 and 81% of the total variation in the soil bacterial community diversity and composition. Overall, the bacterial community composition was strongly driven by the soil moisture, pH, SOC, and diversity (Figure 7).

**DISCUSSION**

In this research, we discovered that short-term N additions and seasonal variations may have a significant impact on soil physical and chemical characteristics, and therefore affect soil bacterial community diversity and structure. Soil acidification was induced by simulated N inputs, and soil moisture, pH, and SOC varied substantially throughout the three monitoring months. These soil characteristics were strongly linked to changes in the soil bacterial community. Furthermore, the reactions of the soil bacterial community to N deposition were seasonally dependent.

**Effects of N Addition on Soil Bacterial Diversity**

In November, the OTU richness and diversity of the soil bacterial community declined linearly with increased N inputs, according to our data. A similar finding was published in the same area, demonstrating that the bacterial diversity of BSCs (i.e., the ACE and Chao1 index) declined linearly with increasing N additions (Wang et al., 2015). Furthermore, our findings revealed that MN and HN addition substantially reduced soil bacterial diversity indices (Shannon and Pielou) in November and March, but had no effect in May. This finding suggested that soil environment variables may influence the impact of N addition on bacterial diversity in the soil. Soil moisture varied considerably across the
three study months, with the lowest in May (11%) and the greatest in November (21%). According to Lucas et al. (2011), excessive soil moisture may enhance soil acidification owing to increased metal cation leaching. N addition did definitely reduce soil pH in this alkaline desert soil, particularly in November (0.35 units) and March (0.38 units), which is approximately one unit more than the stated worldwide average of 0.26 units (Tian et al., 2016). According to the correlation study, soil bacterial diversity indices were positively associated with soil pH ($P < 0.05$; Supplementary Table 3). In our research, there were no reactions of soil bacterial diversity to N addition in May, which may be attributed to minor variations in soil pH (0.08 units). Thus, in desert environments, the differential reactions of soil bacteria to N addition may be strongly related to water availability and variations in soil pH.

**Effects of N Addition on Soil Bacterial Community Composition**

The composition of soil bacterial communities is closely related to ecosystem functions (Allison et al., 2013). Understanding the mechanisms underlying the effects of environmental change on soil bacterial communities will thus aid in forecasting the response of ecosystem processes to environmental change (Jansson and Hofmockel, 2019). In contrast to our second hypothesis, seven bacterial phyla are significantly affected by N addition in May (Figure 3). At the high N treatment, for example, bacterial assemblages classified as Proteobacteria, Bacteroidetes, and Acidobacteria decreased. Similarly, Li Q. et al., 2016 discovered that high N treatment increased the relative abundance of Actinobacteria while decreasing the relative abundance of Proteobacteria. Acidobacteria, Alphaproteobacteria, and Bacteroidetes decreased after N enrichment, according to Zeng et al. (2016). Wu et al. (2019) discovered that high N treatment reduced the relative abundance of Chloroflexi and Planctomycetes while increasing the relative abundance of Proteobacteria, Cyanobacteria, Bacteroidetes, and Verrucomicrobia. Wang et al. (2020) discovered that N inputs increased the relative abundance of Actinobacteria and Chloroflexi while decreasing the relative abundance of Acidobacteria and Verrucomicrobia, which contradicted our findings. Furthermore, the one-way ANOVA revealed that there was a strong shift in bacterial relative abundance in May, with N addition having a significant effect on seven bacterial phyla (Figure 3B). In November and March, however, no bacterial phylum showed significant changes (Figures 3A, C). This finding suggested that investigating the effects of N addition on the soil bacterial community over a single season could be biased. Several studies have found seasonal dynamics in the abundance and community composition of soil bacterial communities (Liu et al., 2018; Wang et al., 2020). In our study, November and March are typically characterized by low temperatures, whereas May is typically characterized by high temperatures and dry conditions. Buckeridge et al. (2010) found that when soil temperatures are close to 0°C, soil microorganisms and the C and N pool have high stability, which may limit the soil bacterial community's response to N addition.
The non-linear response of the relative abundance of these bacterial phyla (Proteobacteria, Actinobacteria, Acidobacteria, Chloroflexi, Cyanobacteria, Planctomycetes, Verrucomicrobia) to N inputs is perhaps the most intriguing finding in our study (Figure 3B). Previous research has shown that soil bacterial composition is linked to soil pH and nutrient availability (potassium, magnesium, calcium, and SOC; Lauber et al., 2009; Li et al., 2019; Zhang et al., 2020). The RDA findings revealed that soil pH, AK, and SOC concentration all had a substantial impact on the bacterial community, confirming that soil pH, AK, and SOC content were all key variables affecting bacterial structure. Indeed, soil available K content may be an essential regulator on Proteobacteria and Cyanobacteria abundance shifts, since both altered accordingly with soil available K content variations. In May, however, we discovered that soil pH and SOC concentration dropped linearly with increased N inputs (Figure 1). This finding showed that the shifts of particular bacterial taxonomic groupings could not be anticipated merely by changes in a single soil parameter, and that the complicated response of soil bacterial communities to N addition might be the consequence of a combination
of many variables. As Wang et al. (2015), non-linear variations in the relative abundance of Actinobacteria and Chloroflexi were caused by phased regulation by soil SOC and pH value.

We also found that Rubrobacter and Geodermatophilus had a greater relative abundance, but their abundances did not vary significantly across various N additions. Rubrobacter genera have a general adaptability to hot and dry conditions (Ferreira et al., 1999; Laiz et al., 2009), and certain Rubrobacter and Geodermatophilus species are known to be extremely radiation-resistant (Yoshinaka et al., 1973; Montero-Calasanz et al., 2015). The features of the two aforementioned genera are critical to the development and persistence of BSCs in dry zones with high UV exposure.

Effects of Seasonal Changes on Bacterial Community Diversity and Composition

Previous research indicated that seasonal community dynamics existed in soil bacterial communities (Rasche et al., 2011; Matulich and Martiny, 2016). In our research, the most diverse soil bacterial community was found in May, followed by March and November. Meanwhile, we discovered that the number of bacterial phyla and genera with relative abundance (more than 1%) was essentially constant throughout seasonal fluctuations, and the bacteria phyla in May were more susceptible to N addition than the other 2 months. Changes in relative abundance induced by N addition may be the key to May's greatest bacterial diversity. Furthermore, prior research indicated that seasonal changes in desert soils are driven by variations in soil moisture and temperature, which are known to be significant drivers of changes in soil bacterial communities (Jia et al., 2021). Our findings showed that soil moisture was strongly linked to the variety of bacterial communities studied (Supplementary Table 3). Lan et al. (2018) also found that increasing soil moisture reduced bacterial community diversity, which is consistent with the findings of this research.

Seasonal changes in the composition of soil bacterial communities may be driven by the shifts in climate and resource availability (Firestone, 2006). The soil in our research location, situated in Xinjiang's Gurbantünggüt Desert, has a relatively high pH and low soil moisture. High pH and low soil moisture were linked to greater evaporation in May when compared to March and November (Yu et al., 2014). Soil moisture, nutrient availability, and pH are examples of environmental factors that may be used to explain the substantial variations in soil microbial community composition between seasons (Smith et al., 2010; Rasche et al., 2011). We discovered that variations in soil nutrient availability may cause the soil bacterial community's (oligotroph-coiotroph) nutritional strategy to alter. Proteobacteria and Bacteroidetes are copiotrophic microorganisms with fast growth among the major bacterial groupings (Fierer et al., 2012, 2007). Both of these soil microbial groups were more abundant in March and November, when soil nutrient concentration was greater than in May. Moreover, these bacteria are gram-negative and lack thick cell walls (Fierer et al., 2007; Schimel et al., 2007). To avoid cell rupture, most gram-negative bacteria can quickly dispose of osmolytes upon rewetting, either via respiration, polymerization, or transfer across the cell membrane. As a result, gram negative bacteria may flourish in the time immediately after rewetting (Schimel et al., 2007). Furthermore, the cold-resistant Bacteroidetes (Flexibacter sp., Sphingobacterium sp., Flavobacterium sp., Flectobacillus sp.) allow members of this phylum to survive in cold settings (Margesin and Schinner, 1994; An et al., 2010). Cyanobacteria, the first group of photosynthetic gram-negative prokaryotes to evolve oxygen, had a greater relative abundance in May (Abed et al., 2010). These results suggest that seasonal variations in bacterial community composition may be ascribed to bacteria's physiological adaptation to soil temperature, precipitation, and vegetation conditions.

CONCLUSION

In November, bacterial OTU richness and diversity declined linearly with increased N inputs, indicating that the impact of N addition on bacterial diversity was seasonally dependent. MN and HN supplementation substantially reduced bacterial diversity indices in November and March (Shannon and Pielou). In May, N addition had a significant impact on the relative abundance of seven bacterium phyla, and the relative abundance of seven bacteria phyla reacted non-linearly to N addition. High N inputs reduced the relative richness of the bacterial phyla Proteobacteria, Bacteroidetes, and Acidobacteria while increasing the number of Actinobacteria and Chloroflexi. Concurrently, the variety and composition of the soil bacterial community changed significantly across months, with bacterial community diversity indices peaking in May. The composition of the soil bacterial community correlated strongly with soil pH, moisture, SOC, and AK concentrations, suggesting that these factors were important in maintaining the stability of the soil bacterial community. Taken together, the reactions of the soil bacterial community to increasing N vary substantially between seasons, and seasonal variations influenced the soil bacterial community through influencing soil characteristics.

DATA AVAILABILITY STATEMENT

The datasets presented in this study can be found in online repositories. The names of the repository/repositories and accession number(s) can be found below: https://dataview.ncbi.nlm.nih.gov/object/PRJNA701371?reviewer=92edmcq2pqqi57f915t3tm9nu.

AUTHOR CONTRIBUTIONS

TH: conceptualization, data curation, formal analysis, investigation, software, validation, visualization, and writing – original draft. WL: conceptualization, funding acquisition, investigation, writing, review, and editing. X-EL: data curation, software, writing, review, and editing. YJ and XW: writing,
review, and editing. YC: funding acquisition, writing, review, and editing. All authors contributed to the article and approved the submitted version.

FUNDING

This work was supported by the Key Laboratory Project, Xinjiang (Grant No. 2021D04006), the National Natural Science Foundation of China (NSFC, Grant No. 32160281), and the Changjiang Scholars Program (111001-40005002).

REFERENCES

Abed, R., Al-Kharusi, S., and Al-Hinai, M. (2015). Effect of biostimulation, temperature and salinity on respiration activities and bacterial community composition in an oil polluted sediment soil. Int. Biodeterior. Biodegrad. 98, 43–52. doi: 10.1016/j.ibiod.2014.11.018

Abed, R., Dobretsov, S., and Sadesh, K. (2010). Applications of cyanobacteria in biotechnology. J. Appl. Microbiol. 106, 1–12.

Allison, S. D., Lu, Y., Weihe, C., Goulden, M. L., Martiny, A. C., Treseder, K. K., et al. (2013). Microbial abundance and composition influence litter decomposition response to environmental change. Ecology 94, 714–725. doi: 10.1890/12-1243.1

An, L., Chen, Y., Xiang, S., Shang, T., and Tian, L. (2010). Differences in community composition of bacteria in four glaciers in Western China. Biogeosciences 7, 1937–1952. doi: 10.5194/bg-7-1937-2010

Angel, R., and Conrad, R. (2013). Elucidating the microbial resuscitation cascade in biological soil crusts following a simulated rain event. Environ. Microbiol. 15, 2799–2815.

Belnap, J., and Lange, O. L. (2001). Biological Soil Crusts: Structure, Function, and Management. Berlin, Germany: Springer-Verlag.

Bensky, D., Gamble, A., and Kaptchuk, T. J. (1993). Chinese herbal medicine: materia medica. Seattle: Eastland Press.

Bobbiuk, R., Hicks, K., Galloway, J., Stroup, T., Alkemade, R., Ashmore, M., et al. (2010). Global assessment of nitrogen deposition effects on terrestrial plant diversity: a synthesis. Ecol. Appl. 20, 30–59. doi: 10.1890/08-1140.1

Buckeridge, K. M., Cen, Y. P., Grogan, L. P., and Layzell, D. B. (2010). Soil biogeochemistry during the early spring in low arctic mesic tundra and the impacts of deepened snow and enhanced nitrogen availability. Biogeochemistry 99, 127–141. doi: 10.1007/s10533-009-9396-7

Caporaso, J. G., Kuczynski, J., Stombaugh, J., Bittinger, K., Bushman, F. D., Costello, E. K., et al. (2010). QIIME allows analysis of high throughput community sequencing data. Nat. Methods 7, 335–336.

Cregger, M. A., Schadt, C. W., Mcdowell, N. G., Pockman, W., and Classen, A. T. (2012). Response of the soil microbial community to changes in precipitation in a semiarid ecosystem. Appl. Environ. Microbiol. 78:8587. doi: 10.1128/aem.02050-12

DeForest, J. L., Zak, D. R., Pregitzer, K. S., and Burton, A. J. (2004). Atmospheric nitrate deposition, microbial community composition, and enzyme activity in northern hardwood forests. Soil Sci. Soc. Am. J. 68, 132–138. doi: 10.2136/sssaj2004.1320

Edgar, R. C., Haas, B. J., Clemente, J. C., Quince, C., and Knight, R. (2011). UCHIME improves sensitivity and speed of chimera detection. Bioinformatics 27, 1994–2000. doi: 10.1093/bioinformatics/btr381

Ferreira, A. C., Nobre, M. F., Moore, E., Rainey, F. A., Battista, J. R., and Costa, M. S. (1999). Characterization and radiation resistance of new isolates of Rubrobacter radiotolerans and Rubrobacter xylanophilus. Extremophiles 3, 235–238. doi: 10.1007/s00793920050121

Fierer, N., Bradford, M. A., and Jackson, R. B. (2007). Toward an ecological classification of soil bacteria. Ecology 88, 1354–1364. doi: 10.1890/05-1839

Fierer, N., Lauber, C. L., Ramirez, K. S., Zaneveld, J., Bradford, M. A., and Knight, R. (2012). Comparative metagenomic, phylogenetic and physiological analyses of soil microbial communities across nitrogen gradients. ISME J. 6, 1007–1017. doi: 10.1038/ismej.2011.159

ACKNOWLEDGMENTS

We would like to thank TopEdit (www.topeditsci.com) for its linguistic assistance during the preparation of this manuscript.

SUPPLEMENTARY MATERIAL

The Supplementary Material for this article can be found online at: https://www.frontiersin.org/articles/10.3389/fmicb.2021.665975/full#supplementary-material

Firestone, M. P. W. K. (2006). Response of microbial community composition and function to soil climate change. Microb. Ecol. 52, 716–724. doi: 10.1007/s00248-006-9103-3

Galloway, J. N., Winiwarter, W., Leip, A., Leach, A. M., Bleeker, A., and Erisman, J. W. (2014). Nitrogen footprints: past, present and future. Environ. Res. Lett. 9:115003. doi: 10.1088/1748-9326/9/11/115003

Huang, G., Li, Y., and Su, Y. G. (2015). Divergent responses of soil microbial communities to water and nitrogen addition in a temperate desert. Geoderma 251, 55–64. doi: 10.1016/j.geoderma.2015.03.018

Hui, R., Zhao, R. M., Liu, L. C., Li, G., Yang, H. T., Gao, Y. H., et al. (2016). Modelling the influence of snowfall on cyanobacterial crusts in the Gurbantunggut Desert, Northern China. Aust. J. Bot. 64, 476–483. doi: 10.1071/bt16008

Janssen, J. K., and Hofmockel, K. S. (2019). Soil microbiomes and climate change. Nat. Rev. Microbiol. 18, 35–46. doi: 10.1038/s41579-019-0265-7

Jia, M., Gao, Z., Gu, H., Zhao, C., Liu, M., Liu, F., et al. (2021). Effects of precipitation change and nitrogen addition on the composition, diversity, and molecular ecological network of soil bacterial communities in a desert steppe. PLoS One 16:e0248194. doi: 10.1371/journal.pone.0248194

Jia, T., Guo, T., Yao, Y., Wang, R., and Chai, B. (2020). Seasonal microbial community characteristic and its driving factors in a copper tailings dam in the Chinese Loess Plateau. Front. Microbiol. 11:1574. doi: 10.3389/fmicb.2020.01574

Juha, M., Alatalo, Annika, K., Jägerbrand, and Ulf, M. (2015). Testing reliability of short-term responses to predict longer-term responses of bryophytes and lichens to environmental change. Ecol. Indic. 58, 77–85.

Kennedy, A. C., and Smith, K. L. (1995). Soil microbial diversity and the sustainability of agricultural soils. Plant Soil 170, 75–86. doi: 10.1007/978-94-011-0479-1_6

Laiz, L., Miller, A. Z., Jurado, V., Akatova, E., Sanchez-Moral, S., Gonzalez, J. M., et al. (2009). Isolation of five Rubrobacter strains from biodeteriorated monuments. Die Naturwissenschaften 96, 71–79. doi: 10.1007/s00114-008-0452-2

Lan, G. Y., Wu, Y., Lesueur, Z. X., and Xie, G. S. (2018). Seasonal changes impact soil bacterial communities in a rubber plantation on Hainan Island, China. Sci. Total Environ. 626, 826–834. doi: 10.1016/j.scitotenv.2018.01.147

Lauber, C. L., Hamady, M., Knight, R., and Fierer, N. (2009). Pyrosequencing-based assessment of soil pH as a predictor of soil bacterial community structure at the continental scale. Appl. Environ. Microbiol. 75, 5111–5120.

LeBauer, D. S., and Treseder, K. K. (2008). Nitrogen limitation of net primary productivity in terrestrial ecosystems is globally distributed. Ecology 89, 371–379. doi: 10.1890/06-2057.1

Li, H., Xu, Z., Yang, S., Li, X. P., Top, E. M., Wang, R. Z., et al. (2016). Responses of soil bacterial communities to nitrogen deposition and precipitation increment are closely linked with aboveground community variation. Microb. Ecol. 71, 974–989. doi: 10.1007/s00248-016-0730-z

Li, Q., Song, X. Z., Gu, H. H., and Gao, F. (2016). Nitrogen deposition and management practices increase soil microbial biomass carbon but decrease diversity in Moso bamboo plantations. Sci. Rep. 6:28235.

Li, K. H., Liu, X. J., Song, W., Chang, Y. H., and Hu, Y. K. (2014). Atmospheric nitrogen deposition at two sites in an arid environment of Central Asia. PLoS One 9:e007018. doi: 10.1371/journal.pone.007018
Nitrogen Alter Soil Bacterial Community

Wang, J., Shi, X., Zheng, C., Suter, H., and Huang, Z. (2020). Different responses of soil bacterial and fungal communities to nitrogen deposition in a subtropical forest. Sci. Total Environ. 755:142449. doi: 10.1016/j.scitotenv.2020.142449

Wang, Q., Garrity, G. M., Tiedje, J. M., and Cole, J. R. (2007). Naïve Bayesian classifier for rapid assignment of rRNA sequences into the new bacterial taxonomy. Appl. Environ. Microbiol. 73, 5261–5267. doi: 10.1128/aecm.00626-07

Wu, J., Liu, W., Zhang, W., and Fan, H. (2019). Long-term nitrogen addition changes soil microbial community and litter decomposition rate in a subtropical forest. Appl. Soil Ecol. 142, 43–51. doi: 10.1016/j.apsoil.2019.05.014

Wu, K., Xu, W. X., and Yang, W. Y. (2020). Effects of precipitation changes on soil bacterial community composition and diversity in the Junggar Desert of Xinjiang, China. PeerJ. 8:e8433. doi: 10.7717/peerj.8433

Xia, X., Guo, W., and Liu, H. (2015). Dynamics of the bacterial and archaeal communities in the Northern South China Sea revealed by 454 pyrosequencing of the 16S rRNA gene. Deep Sea Res. Part II Top. Stud. Oceanogr. 117, 97–107. doi: 10.1016/j.dsr2.2015.05.016

Yang, Y., Chen, H., Gao, H., and An, S. (2020). Response and driving factors of soil microbial diversity related to global nitrogen addition. Land Degradat. Dev. 6, 1–15.

Yoshinaka, T., Yano, K., and Yamaguchi, H. (1973). Isolation of a highly radioresistant bacterium. Arthrobacter radiotolerans nov.sp. Agric. Biol. Chem. 37, 2269–2275. doi: 10.1271/bbb1961.37.2269

Yu, J., Glazer, N., and Steinberger, Y. (2014). Carbon utilization, microbial biomass, and respiration in biological soil crusts in the Negev Desert. Biol. Fertil. Soil 50, 285–293. doi: 10.1007/s00374-013-0856-9

Yue, P., Zuo, X., Li, K., Gui, X., and Liu, X. (2021). The driving effect of nitrogen-related functional microorganisms under water and nitrogen addition on N2O emission in a temperate desert. Sci. Total Environ. 772:145470. doi: 10.1016/j.scitotenv.2021.145470

Zeng, J., Liu, X., Song, L., Lin, X., Zhang, H., Shen, C., et al. (2016). Nitrogen fertilization directly affects soil bacterial diversity and indirectly affects bacterial community composition. Soil Biochem. Biochem. 92, 41–49. doi: 10.1016/j.soilbio.2015.09.018

Zhang, Y.Q., Hou, L. Y., Li, Z. C., Zhao, D. X., Song, L. G., Shao, G. D., et al. (2020). Leguminous supplementation increases the resilience of soil microbial community and nutrients in Chinese fir plantations. Sci. Total Environ. 703, 134917:1–134917:11.

Zhao, X., Zhang, Y., and Downing, A. (2012). Non-linear response of microbial activity across a gradient of nitrogen addition to a soil from the Gurbantungut Desert, Northwestern China. Soil Biochem. Biochem. 47, 67–77. doi: 10.1016/j.soilbio.2011.05.012

Conflict of Interest: The authors declare that the research was conducted in the absence of any commercial or financial relationships that could be construed as a potential conflict of interest.

Publisher's Note: All claims expressed in this article are solely those of the authors and do not necessarily represent those of their affiliated organizations, or those of the publisher, the editors and the reviewers. Any product that may be evaluated in this article, or claim that may be made by its manufacturer, is not guaranteed or endorsed by the publisher.

Copyright © 2021 Huang, Liu, Long, Jia, Wang and Chen. This is an open-access article distributed under the terms of the Creative Commons Attribution License (CC BY). The use, distribution or reproduction in other forums is permitted, provided the original author(s) and the copyright owner(s) are credited and that the original publication in this journal is cited, in accordance with accepted academic practice. No use, distribution or reproduction is permitted which does not comply with these terms.