N/Fe/Zn co-doped TiO$_2$ loaded on basalt fiber with enhanced photocatalytic activity for organic pollutant degradation

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To avoid the loss of catalytic material powder, a loaded catalytic material of TiO$_2$ with basalt fiber as the carrier (TiO$_2$@BF) was synthesized by an improved sol–gel method. The TiO$_2$@BF was doped with different contents of N, Fe and Zn elements and was used to degrade rhodamine B (RhB) under ultraviolet light. The physical characterization analysis indicated that the co-doping of the N, Fe and Zn elements had the effects of reducing grain size, increasing sample surface area, and narrowing the electronic band gap. The electronic band gap of nitrogen–iron–zinc co-doped TiO$_2$@BF (N/Fe/Zn$_x$TiO$_2$@BF) was 2.80 eV, which was narrower than that of TiO$_2$@BF (3.11 eV). The degradation efficiency of RhB with N/Fe/Zn$_x$TiO$_2$@BF as a photocatalyst was 4.3 times that of TiO$_2$@BF and its photocatalytic reaction was a first-order kinetic reaction. Quenching experiments suggested that the reactive species mainly include photoinduced holes (h$^+$), superoxide radicals (‘O$_2^-$) and hydroxyl radicals (’OH). In brief, this study provides a prospective loaded catalytic material and routine for the degradation of organic contaminants in water by a photocatalytic process.

1 Introduction

At present, water pollution caused by man-made organic pollutants has become one of the key concerns all over the world. Large quantities of organic compounds, such as synthetic dyes and endocrine disruptors, originating from industries including textiles, plastics, pharmaceuticals, etc., would have harmful effects on mankind and animal life in short or long periods of time. Water pollution caused by dye waste-water discharged from traditional industries, one of the most common contaminants, has attracted extensive attention. Rhodamine B (RhB), a typical pollutant of dye molecules, is difficult to be removed by self-purification of natural water systems because of its high stability and it greatly threatens human health due to its toxicity. Therefore, research into removing dye molecules in water has become a hot topic in recent years.

Nowadays, common water treatment methods mainly include membrane filtration, coagulation sedimentation, electrochemical, advanced oxidation process, and biological treatment methods. Photocatalytic oxidation, an advanced oxidation process, could efficiently eliminate various harmful organic compounds. It has attracted extensive attention in the field of water environmental protection and restoration due to the advantages of low cost, easy operation, high efficiency, low consumption, no secondary pollution, and great damage to dye molecules.

In the study of photocatalytic oxidation to degrade pollutants, TiO$_2$ nanoparticles are one of the most widely studied semiconductor photocatalysts in photocatalysis due to their strong oxidizing ability, chemical stability, safety, non-toxicity, and high photocatalytic potential. However, TiO$_2$ has some intrinsic drawbacks in practical applications such as its wide band gap and charge carrier recombination, which result in low light utilization efficiency and photocatalytic activity in the visible region. Consequently, tremendous efforts have been devoted to improving the photocatalytic efficiency of TiO$_2$ via modification, including doping modification, surface modification, dye sensitization, deposition of noble metals, semiconductor recombination, etc. In particular, metal doping and non-metal doping for TiO$_2$ have proved to be the most effective methods and have been extensively studied.

In addition, TiO$_2$ nanoparticles are difficult to recycle and reuse in applications because of their small particles. In order to solve the problem of catalyst recycling and reuse, many researchers have used different substances as carriers to support TiO$_2$ to synthesize TiO$_2$ loaded composites, such as activated carbon, glass fiber and carbon fiber.
(BF) is made of pure natural basalt ore as the raw material and its main component is SiO$_2$, which has a similar structure to TiO$_2$. Therefore, BF, which has similar mineral properties to TiO$_2$, could adhere to TiO$_2$ firmly. Generally, to synthesize TiO$_2$ loaded composites, TiO$_2$ nanoparticles need to be prepared first and then dispersed in a sol to carry out loading. The high temperature resistance and excellent physicochemical stability of BF allows it to be loaded in the intermediate step of TiO$_2$ preparation and then calcined to obtain the composite material in one step, thereby increasing the loading efficiency of TiO$_2$ and saving time and cost. To the best of our knowledge, no studies on TiO$_2$ loaded on BF for the degradation of organic contaminants in water have been reported.

In the current study, a new three-dimensional structure model was made from BF and TiO$_2$ with three doping elements of N, Fe, and Zn loaded on the BF to prepare a multi-element co-doped TiO$_2$@BF composite material, which has a larger contact area with water and a higher light transmittance. The synthesized new multi-element doped G$_x$TiO$_2$@BF (in G$_x$TiO$_2$@BF, G represents the dopants N, Fe and Zn) composite material was characterized. Meanwhile, the doping substance combination and the effects of the dopant content (%) of different dopants were investigated, the photocatalytic performance of the synthesized material was tested for the degradation of RhB and the possible degradation mechanisms were studied.

2 Experimental materials and methods

2.1 Chemical reagents

Tetrabutyl titanate ($\geq$98.5%), ethyl alcohol ($\geq$99.5%, anhydrous), and nitric acid (65–68%) were used for TiO$_2$ synthesis. Urea ($\geq$99.0%), ZnO ($\geq$98.0%), and ferric nitrate nonahydrate ($\geq$98.5%) were used as reagents when doping and modifying TiO$_2$. Rhodamine B (RhB) was used in the photocatalytic degradation experiment. 1,4-Benzquinone (BQ, $\geq$98.5%), tert-butyl alcohol (TBA, $\geq$99.0%) and EDTA-2Na ($\geq$99.0%) were used in the reactive species quenching experiments. All the reagents were from Chron Chemicals Co., Ltd. (Sichuan, China). Distilled water (resistance $> 18.25$ M$\Omega$), used in all experiments, was provided by a water purification system (ULUPURE, China).

2.2 Synthesis of the catalytic loaded composites

2.2.1 Colloidal sol. Typically, 200 mL of tetrabutyl titanate was added dropwise to a 1000 mL sized beaker containing 600 mL ethyl alcohol with constant mechanical stirring. A mixture, 30 mL of nitric acid was added to the mixture slowly and uniformly, then the beaker was sealed and stirring continued for 1 hour. Finally, a yellow transparent sol was obtained.

Fig. 1 Preparation steps of G$_x$TiO$_2$@BF and the photocatalytic degradation reaction device.
2.2.2 **Doping.** N\_TiO\_2 sol, Fe\_TiO\_2 sol and Zn\_TiO\_2 sol were synthesized, respectively, doping substances of urea, ferric nitrate nonahydrate and ZnO with different doping ratios. The corresponding proportions were \( n(N) : n(Ti) = (6\%, 8\%, \text{and} 12\%) \), \( n(Fe) : n(Ti) = (0.03\%, 0.3\%, \text{and} 1\%) \), and \( n(Zn) : n(Ti) = (0.5\%, 2\%, \text{and} 7\%) \). The doping substances with different doping ratios were separately dissolved in 600 mL ethyl alcohol and then the mixed solution was used as an organic solvent. The remaining steps were consistent with the preparation of the sol. After the photocatalytic degradation of RhB with single doping of TiO\_2@BF, the optimal doping ratio was selected to be 8\% N, 0.3\% Fe and 2\% Zn, and then the three materials were co-doped with an optimal doping ratio of N, Fe, and Zn to obtain the N/Fe/ Zn\_TiO\_2 sol.

2.2.3 **Catalytic loaded composites.** After the sol was prepared, the three-dimensional model of BF shown in Fig. 1 was evenly loaded with a layer of sol by the dipping method and the second impregnation was carried out after the first layer was naturally air-dried. After that, the BF model was placed in a drying oven at a temperature of 85 °C for 50 minutes to volatilize the ethyl alcohol, then put in a muffle furnace at 400 °C and calcined for 3 hours. Finally, TiO\_2@BF, N\_TiO\_2@BF, Fe\_TiO\_2@BF, Zn\_TiO\_2@BF and N/Fe/Zn\_TiO\_2@BF were obtained. The TiO\_2 loading quantity was calculated by the mass difference between the three-dimensional model of BF without loading TiO\_2 and TiO\_2@BF.

Fig. 1 presents the preparation steps of loading TiO\_2 on BF and the photocatalytic degradation reaction device. The BF model was based on a round wire mesh as a fiber guide bracket; three equal-height wires were used as a wire mesh fixed bracket and 126 basalt fiber bundles were fixed at 5 mm intervals on it. Every basalt fiber bundle was 27 cm in length and 0.5 mm in diameter.

2.3 **Characterization**

The surface morphology and fracture morphology of the sample were examined by scanning electron microscopy (SEM, S-450, Japan). Energy dispersive spectrometry (EDS, Aztec X-Max80, UK) was applied to the qualitative analysis of elements. The crystalline phases and grain size of the samples were characterized by X-ray diffraction (XRD, EMPYREAN, UK), using Cu K\(\alpha\) radiation (\(\lambda_1 = 1.5406 \text{ Å}, \lambda_2 = 1.5444 \text{ Å}\)), a monochromatic slit of 0.76 mm, a 2\(\theta\) diffraction angle of 10° to 80° and a scan step of 0.03° s\(^{-1}\). The functional groups were analyzed by Fourier transform infrared spectroscopy (FT-IR, Nicolet 6700, USA). The molecular structural groups were analyzed by Raman spectra (RAMAN, LabRAM HR800, France). The surface composition and chemical state of the sample were investigated by the AXIS

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**Fig. 2**  SEM micrographs for the samples: (a) BF and (b) TiO\_2@BF; EDS elemental mapping images of (c)–(h) N/Fe/Zn\_TiO\_2@BF.
Supra multifunctional X-ray photoelectron spectrometer (XPS, KRATOS, UK), using a monochromatic X-ray source. The electronic band gaps of the synthesized photocatalysts were analyzed by UV-visible diffuse-reflectance spectroscopy (UV-vis/DRS, UV-3600, Japan). The samples used for characterization like TiO$_2$ refer to powders scratched off from the fiber and those like TiO$_2$@BF refer to layers loaded on basalt fibers.

2.4 Photocatalytic degradation of RhB

The photocatalytic degradation of RhB was investigated under ultraviolet (UV) light in a photocatalytic dark box reactor. The experiments were performed in a 10 L cylindrical black box with a mercury lamp (42 W, $\lambda = 253.7$ nm) as the light source, and circulating water was adopted to control the reaction temperature at ambient temperature. In order to eliminate the disturbance of the adsorption, it is necessary to reach an adsorption/desorption equilibrium before turning on the light.

The stability and reusability experiment was obtained at high-speed centrifugation before detecting the RhB concentration by spectrophotometry (V729, YOKE INSTRUMENT, China) at 554 nm. The degree of RhB degradation was expressed as $\text{C}_0/\text{C}_t$, where $C_0$ was the RhB concentration at the sampling time, and $C_t$ was RhB concentration at the initial time. The stability and reusability experiment materials were washed with absolute ethanol and deionized water several times. The samples were calcined at 400 °C for 1 h before the next cycle experiment was carried out.

3 Results and discussion

3.1 Physicochemical properties

3.1.1 Crystal structure, crystallite size and morphology.

The photocatalytic activity is greatly influenced by the crystal phase of TiO$_2$. As reported, the photocatalytic activity of anatase phase TiO$_2$ synthesized at 400 °C is better than that of the rutile phase and brookite phase. Therefore, 400 °C was selected as the calcination temperature to prepare the catalyst.

Fig. 2(a) and (b) depict the microstructure and morphology of BF and TiO$_2$@BF by SEM, from which it could be obviously observed that the fibers were wrapped with TiO$_2$. There were many micropores on the surface of the synthesized N/Fe/Zn_TiO$_2$@BF catalyst due to using urea as the N doping source. During the calcination process, urea could change to gas and escape, increasing the specific surface area of material. As shown in Fig. 2(c)-(h), it was confirmed that the O, Ti, N, Fe and Zn elements coexisted on the N/Fe/Zn_TiO$_2$@BF.

XRD is performed to characterize the crystal structure of the material. TiO$_2$ was stripped from the BF surface to analyze its crystal phase due to the diffraction peak of BF covering that of TiO$_2$. Fig. 3 depicts the XRD results of the synthesized TiO$_2$ and N/Fe/Zn_TiO$_2$ and TiO$_2$@BF. The diffraction peak of synthesized TiO$_2$ appeared at $2\theta = 25.22^\circ$ (101), $37.89^\circ$ (004), $48.07^\circ$ (200), $53.81^\circ$ (105), $55.08^\circ$ (211) and $62.63^\circ$ (204), which matched well with the JCPDS NO: 21-1272. The result shows that the synthesized TiO$_2$ was monocrystalline and in the pure anatase phase, which was confirmed by there being no diffraction peaks at $2\theta = 27.5^\circ$ (110) and $30.8^\circ$ (121), corresponding to the rutile and brookite phase. The diffraction peak location of N/Fe/Zn_TiO$_2$ also corresponded to the anatase phase, indicating that the doping of N, Fe and Zn did not change the crystal structure of pristine TiO$_2$. In addition, the XRD spectra of N/Fe/Zn_TiO$_2$ (Fig. 3) show no peaks of Zn oxide and Fe oxide.

The crystallite sizes were calculated from the stretching of the anatase (101) peaks by using the Scherrer formula as follows.

$$D = \frac{k\lambda}{\beta \cos \theta}$$

where $D$ is the crystallite size (nm), $k$ is the Scherrer constant (0.89), $\lambda$ is the wavelength of X-ray (1.5406 Å), $\beta$ is the half-height width of the diffraction peak of the measured sample (rad), and $\theta$ is the Bragg diffraction angle (°).

As shown in Fig. 3, TiO$_2$@BF covering that of TiO$_2$ in the RAMAN spectra of the synthesized TiO$_2$. As displayed in Fig. 4(b), their TiO$_2$ crystals were both in the anatase phases. As could be seen from Fig. 4(b), the positions of the characteristic bands in the synthesized TiO$_2$ of 143.2, 197, 395.4, 515 and 639 cm$^{-1}$ correspond to the vibration modes of $E_g$, $E_g$, $B_{1g}$, $A_{1g}$ and $E_g$ in
the single crystal anatase phase, respectively. The result that only five vibration modes appeared was lower than the six Raman activation modes \((A_{1g} + 2B_{1g} + 3E_g)\) reported by Ohsaka\(^{34}\) because the peaks at positions 513 and 515 cm\(^{-1}\) overlap; a similar result was also obtained by Choi\(^{35}\) and Cho.\(^{36}\) In Fig. 4(b), no characteristic bands were observed at 445 and 612 cm\(^{-1}\), indicating that the rutile phase did not exist in the prepared TiO\(_2\) sample. Compared with the characteristic peaks of TiO\(_2\), the characteristic peaks of N/Fe/Zn_TiO\(_2\) were weakened and the half-peak width increased due to the decrease of crystallinity size after doping, which was consistent with the XRD results.

3.1.2 Chemical state, electronic structure and electronic band gaps. The chemical composition of the surface of TiO\(_2\)@BF and N/Fe/Zn_TiO\(_2\)@BF was determined by XPS, as shown in Fig. 5. Compared with the full XPS spectrum of TiO\(_2\)@BF (Fig. 5(a)), the characteristic peaks of Zn and N appeared in the full XPS spectrum of N/Fe/Zn_TiO\(_2\)@BF (Fig. 5(b)) but the Fe signal was not detected due to the low content of Fe in N/Fe/Zn_TiO\(_2\)@BF. The Zn 2p core spectrum in Fig. 5(c) shows that the peaks of 1045.08 eV and 1021.97 eV correspond to Zn 2p\(_{1/2}\) and Zn 2p\(_{3/2}\), respectively, indicating that Zn in N/Fe/Zn_TiO\(_2\)@BF had +2 valence.\(^{37}\) In Fig. 5(d), no characteristic bands were observed at 445 and 612 cm\(^{-1}\), indicating that the rutile phase did not exist in the prepared TiO\(_2\) sample. A similar result was also obtained by Choi\(^{35}\) and Cho.\(^{36}\) In Fig. 5(e), no characteristic bands were observed at 445 and 612 cm\(^{-1}\), indicating that the rutile phase did not exist in the prepared TiO\(_2\) sample. Compared with the characteristic peaks of TiO\(_2\), the characteristic peaks of N/Fe/Zn_TiO\(_2\) were weakened and the half-peak width increased due to the decrease of crystallinity size after doping, which was consistent with the XRD results.

\[
\begin{align*}
C_x &= \frac{A_x}{S_x} \left( \frac{\Sigma A_i}{S_i} \right) \\
&= \text{(the concentration of the element at the surface,} A_x = \text{the XPS peak area of the desired element,} S_x = \text{the relative sensitivity factor of the XPS peak,} \Sigma A_i/S_i = \text{the sum of the ratio of the peak area to the relative sensitivity factor of all the measured elements. The XPS peak area values of N 1s, Zn 2p and Ti 2p were 141.55, 2466.49 and 26432.85, obtained from the XPS spectrum of N/Fe/Zn_TiO\(_2\)@BF. When the concentration ratio was obtained, the molar ratio could be further calculated to be} n(N) : n(Ti) = 8\% \text{ and } n(Zn) : n(Ti) = 2\%. \text{The Fe peak was not detected due to the low Fe content in N/Fe/Zn_TiO\(_2\)@BF. The real doping content of N/Fe/Zn_TiO\(_2\)@BF obtained through the XPS analysis was the same as the theoretical doping content.}
\end{align*}
\]

The FT-IR spectra of the synthesized TiO\(_2\) and N/Fe/Zn_TiO\(_2\) are shown in Fig. 6, which was used to analyze the functional groups of the samples. In Fig. 6(a), the peaks at 3439 cm\(^{-1}\) and 1630 cm\(^{-1}\) were formed by the stretching vibration of the water adsorbed on the surface, \('OH or the H–OH bond of the COOH group, and the peak at 513 cm\(^{-1}\) was attributed to the characteristic Ti–O bond vibration of TiO\(_2\),\(^{35,36}\) In Fig. 6(b), compared with Fig. 6(a), there was a slight offset value at the characteristic peak of TiO\(_2\) and the absorption peak of the hydroxy group. Combined with the results of the XRD and XPS, it indicated that N, Fe and Zn entered into the lattice of TiO\(_2\).

3.2 Photocatalytic degradation of RhB

As shown in Fig. 7(a), RhB removal by TiO\(_2\) (17.93\%) alone was a little better than that of TiO\(_2\)@BF (15.61\%), which was due to the TiO\(_2\) in TiO\(_2\)@BF being wrapped, leading to a lower actual amount of the catalyst working in the photocatalysis.

The photocatalytic activities of three different proportions of TiO\(_2\)@BF doped with N (N_TiO\(_2\)@BF), Fe (Fe_TiO\(_2\)@BF) and Zn (Zn_TiO\(_2\)@BF) were evaluated as shown in Fig. 7(b–d), which show that when the doping ratios of N, Fe and Zn were 8\%, 0.3\% and 2\%, respectively, the degradation effects of RhB under UV light for 3 hours were the best, reaching 47.40\%, 48.20\% and 60.45%, respectively. The reason why the
Fig. 5  XPS surface analysis for TiO$_2$@BF and N/Fe/Zn$_2$TiO$_2$@BF: (a) TiO$_2$@BF full XPS spectrum, (b) N/Fe/Zn$_2$TiO$_2$@BF full XPS spectrum, (c) Zn 2p core level spectra of N/Fe/Zn$_2$TiO$_2$@BF, (d) N 1s core level spectra of N/Fe/Zn$_2$TiO$_2$@BF, (e) Ti 2p core level spectra of TiO$_2$@BF, (f) Ti 2p core level spectra of N/Fe/Zn$_2$TiO$_2$@BF, (g) O 1s core level spectra of TiO$_2$@BF, and (h) O 1s core level spectra of N/Fe/Zn$_2$TiO$_2$@BF.



degradation efficiency of the 8% N, 0.3% Fe and 2% Zn doping ratios was higher than the 12% N, 1% Fe and 7% Zn doping ratios, respectively, might be that when the doping concentration was too high, the doping substance could not easily enter the TiO\(_2\) lattice but instead gathered on the TiO\(_2\) surface, reducing the active sites.

Fig. 7(e) illustrates that the adsorption/desorption equilibrium was reached in 30 min. As shown in Fig. 7(f), the photocatalytic activities of N\(_{\text{TiO}_2}\@BF\), Fe\(_{\text{TiO}_2}\@BF\), and Zn\(_{\text{TiO}_2}\@BF\) on RhB degradation were all better than that of TiO\(_2\@BF\) under UV light. In particular, the degradation efficiency of RhB by N/Fe/Zn\(_{\text{TiO}_2}\@BF\) as the photocatalytic material was the best (67.67%), 4.33 times that of TiO\(_2\@BF\), indicating that N, Fe and Zn element doping enhanced the photocatalytic degradation effect, which was consistent with the UV-vis/DRS characterization analysis results.

Fig. 7(g) shows that the photocatalytic degradation of RhB by 8% N\(_{\text{TiO}_2}\@BF\), 0.3% Fe\(_{\text{TiO}_2}\@BF\), 2% Zn\(_{\text{TiO}_2}\@BF\) and N/Fe/Zn\(_{\text{TiO}_2}\@BF\) conformed to a quasi-first-order kinetic equation. The degradation rate constant \((k)\) of RhB in different processes and the TiO\(_2\) loading amounts for each sample are listed in Table 1. It was found that \(k_{\text{N/Fe/Zn}_{\text{TiO}_2}@BF} < k_{\text{Fe}_{\text{TiO}_2}@BF} < k_{\text{Zn}_{\text{TiO}_2}@BF} < k_{\text{N}_{\text{TiO}_2}@BF}\), and the \(k\) value of one element dopant of TiO\(_2\@BF\) was greater than the \(k\) value of TiO\(_2\@BF\) but less than the \(k\) value of N/Fe/Zn\(_{\text{TiO}_2}\@BF\). The degradation rate constant of N/Fe/Zn\(_{\text{TiO}_2}\@BF\) was 7.04 times that of TiO\(_2\@BF\), indicating that doping indeed accelerated charge transfer and effectively improved the photocatalytic activity. The analysis result of XRD and RAMAN that the crystallite sizes of N/Fe/Zn\(_{\text{TiO}_2}\) were smaller than TiO\(_2\) was consistent with this photocatalytic performance.

In order to estimate the reusability and stability of N/Fe/Zn\(_{\text{TiO}_2}\@BF\) for RhB removal, three recycling tests were conducted at the same experimental conditions. As shown in Fig. 7(h), the degradation effects of RhB were 67.67%, 55.70% and 55.84% in three consecutive cycle experiments, suggesting that N/Fe/Zn\(_{\text{TiO}_2}\@BF\) could maintain relative stability after the first cycle experiment.

### 3.3 Degradation mechanism analysis

#### 3.3.1 Optical characteristics and electronic band gaps analysis of the synthesized photocatalyst

As shown in Fig. 8, the optical characteristics and electronic band gaps of TiO\(_2\) and G\(_{\text{TiO}_2}\) were evaluated by using UV-vis diffuse reflectance. There was an obvious red shift of the absorption band for the N/Fe/Zn\(_{\text{TiO}_2}\) photocatalyst compared to TiO\(_2\). From the Tauc plot using the Kubelka–Munk function (Fig. 8 inset), the electronic band gaps of the synthesized photocatalysts were in the following order: N/Fe/Zn\(_{\text{TiO}_2}\) (2.80 eV) < N\(_{\text{TiO}_2}\) (2.85 eV) < Zn\(_{\text{TiO}_2}\) (3.03 eV) < Fe\(_{\text{TiO}_2}\) (3.08 eV) < TiO\(_2\) (3.11 eV). As a result, N/Fe/Zn\(_{\text{TiO}_2}\) showed the narrowest energy band gap (2.8 eV) compared to the other synthesized photocatalysts, which meant it was easier to generate electronic transitions under the excitation of UV light, indicating that co-doping of the N, Fe, and Zn elements in this study could have a synergistic effect and improve the photocatalytic activity of N/Fe/Zn\(_{\text{TiO}_2}\).

#### 3.3.2 Reactive species quenching experiments

To better elucidate the mechanism of N/Fe/Zn\(_{\text{TiO}_2}\@BF\) for the degradation of rhodamine B, different reactive species quenching experiments were systemically conducted, and the results are presented in Fig. 9. In these experiments, 1,4-benzoquinone (BQ), tert-butyl alcohol (TBA) and EDTA-2Na were used as scavengers, which had different affinities for different reactive oxygen species, such as the active superoxide radical (‘O\(_2^-\)’), hydroxyl radicals (‘OH) and holes (‘h\(^+\)’). It was observed that the degradation effect of RhB (8.33%) was obviously inhibited by the EDTA-2Na scavenger, which indicated that ‘h\(^+\)’ played a crucial role in this photocatalytic degradation process. When BQ and TBA were added, respectively, the degradation rate of RhB also decreased, suggesting that ‘OH and ‘O\(_2^-\)’ have an impact on RhB removal. Herein, RhB removal during the treatment process with the QB scavenger was different from its expected degradation trend. The reason for this result was that some dark colored degradation products of BQ had a positive effect on the determination of RhB by the spectrophotometer method, causing the measured absorbance value to become higher (Fig. 9 inset).

According to the results of the reactive species quenching experiments, it could be inferred that the ‘h\(^+\)’ active species played a dominant role, while ‘O\(_2^-\)’ and ‘OH were the minor radicals in this photocatalytic system. When N/Fe/Zn\(_{\text{TiO}_2}\@BF\) was illuminated by UV light, electrons (e\(^-\)) in the valence band (VB) were excited to the conduction band (CB), leaving holes (‘h\(^+\)’) in the VB, which could directly participate in the RhB degradation (eqn (3) and (4)). Meanwhile, the accumulated e\(^-\) could capture O\(_2\) to obtain ‘O\(_2^-\)’ (eqn (5)), which could capture H\(^+\) generated by the ionization of H\(_2\)O to form H\(_2\)O\(_2\), and the generated ‘OH might cause the translation of the H\(_2\)O\(_2\) (eqn (7)–(9)). The ‘O\(_2^-\)’ and ‘OH radicals are capable of destroying organic compounds (eqn (6) and (10)).

\[
\text{N/Fe/Zn}_{\text{TiO}_2}@BF + hν \rightarrow e^- + h^+ \quad \text{(3)}
\]

\[
h^+ + \text{RhB} \rightarrow \text{degradation} \quad \text{(4)}
\]

\[
e^- + O_2 \rightarrow \text{‘O}_2^- \quad \text{(5)}
\]
Fig. 7  (a) Photocatalytic activity of the same amount of TiO$_2$ without loading on BF and TiO$_2$@BF, (b–d, f) photocatalytic activities (in the degradation of RhB) of TiO$_2$@BF and G$_x$TiO$_2$@BF doped with different substances at different doping ratios, (e) dark adsorption curves, (g) photocatalytic degradation kinetics and (h) the stability and reusability tests of N/Fe/Zn$_x$TiO$_2$@BF for RhB removal.
Catalyst degradation pathways:

\[ \text{CO}_2 + \text{RhB} \rightarrow \text{degradation} \quad (6) \]
\[ \text{H}_2\text{O} \rightarrow \text{H}^+ + \text{OH}^- \quad (7) \]
\[ \text{O}_2^- + 2\text{H}^+ + 2e^- \rightarrow \text{H}_2\text{O}_2 \quad (8) \]
\[ \text{H}_2\text{O}_2 + e^- \rightarrow \text{OH}^- + \text{OH} \quad (9) \]
\[ \text{OH} + \text{RhB} \rightarrow \text{degradation} \quad (10) \]

4 Conclusions

The N/Fe/Zn_TiO_2@BF composites were synthesized by a combined sol-gel calcination method. N_TiO_2@BF, Fe_TiO_2@BF, Zn_TiO_2@BF and N/Fe/Zn_TiO_2@BF exhibited higher photocatalytic activity towards the degradation of RhB under UV light irradiation compared to TiO_2@BF. The degradation efficiency for RhB by N/Fe/Zn_TiO_2@BF was 4.3 times that of TiO_2@BF. The degradation process of N/Fe/Zn_TiO_2@BF was consistent with first-order kinetics with a $k$ value 7.04 times that of TiO_2@BF. The improved photoactivity was attributed to the synergistic effect of the N, Fe and Zn co-doping. The crystallite size of N/Fe/Zn_TiO_2 was only 10.4 nm, while that of TiO_2 without doping was 18.1 nm. There were more catalytic activity points with the specific surface area of N/Fe/Zn_TiO_2@BF increasing greatly, resulting in the improved catalytic efficiency. The results of the EDS, XRD, RAMAN, XPS and FT-IR analysis indicated that the elements of N, Fe and Zn were doped into the crystal lattice of TiO_2 without changing the anatase crystalline phase. In addition, N/Fe/Zn_TiO_2@BF gained a narrower energy band gap (2.80 eV), which improved the photocatalytic activity. Reactive species quenching experiments indicated that h$^+$ was the most important reactive species and ‘O$^\cdot$ played minor roles for the N/Fe/Zn_TiO_2@BF photocatalysis processes in the degradation of RhB.

Conflicts of interest

The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

Table 1: The photo-degradation rate constant and TiO_2 loading amounts of different samples

| Sample             | Doping ratio | $k \times 10^{-3}$ (min$^{-1}$) | TiO_2 loading amount (mg cm$^{-2}$) |
|--------------------|--------------|---------------------------------|------------------------------------|
| TiO_2@BF           | —            | 0.95                            | 3.02                               |
| N_TiO_2@BF         | 6.00%        | 3.07                            | 3.11                               |
| 8.00%              | 3.59         | 3.08                            |
| 12.00%             | 2.50         | 3.08                            |
| Fe_TiO_2@BF        | 0.03%        | 2.63                            | 3.02                               |
| 0.30%              | 3.54         | 3.02                            |
| 1.00%              | 2.34         | 3.06                            |
| Zn_TiO_2@BF        | 0.50%        | 2.73                            | 3.13                               |
| 2.00%              | 4.86         | 3.16                            |
| 7.00%              | 1.33         | 3.08                            |
| N/Fe/Zn_TiO_2@BF   | —            | 6.69                            | 3.04                               |
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