Incorporating adaptation and resilience into an integrated watershed and coral reef management plan

David A. Gibbs, Jordan M. West, Patricia Bradley

1 Oak Ridge Institute for Science Education fellow at U.S. Environmental Protection Agency, Washington, D.C., United States of America, 2 Office of Research and Development, U.S. Environmental Protection Agency, Washington, D.C., United States of America, 3 Patricia Bradley, Tetra Tech, Inc., Owings Mills, MD, United States of America

* Current address: World Resources Institute, Washington, D.C., United States of America
* david.gibbs@wri.org

Abstract

Changing environmental conditions are forcing natural resource managers and communities to adapt their strategies to account for global shifts in precipitation, temperature, sea level and more, all of which are occurring in addition to local human impacts. Adapting to threats from climate change requires a fundamental shift in the practice of natural resource management through the development of forward-looking “climate-smart” goals and strategies. Here we present a proof-of-concept application of a decision-support tool to help design climate-smart management actions for the watershed and coral reef management plan for Guánica Bay watershed in southwest Puerto Rico. We also explore the connection between adaptation planning and coral reef resilience, using a recently developed Puerto Rico-wide reef resilience assessment. In the first phase of the study, we used the publicly available Adaptation Design Tool to draft initial climate-smart versions of twelve proposed management actions. In the second phase, two actions (dirt road management on steep slopes, and coral reef restoration) were further refined through consultations with local experts to make more detailed design adjustments; this included the option to use information from the coral reef resilience assessment to inform design improvements. The first phase resulted in moderately detailed assessments that broadly accounted for anticipated direct and indirect effects of climate change on the planned management actions. The second phase resulted in more site-specific technical assessments and additional important design details. The expert panel charged with discussing climate-smart reef restoration around Guánica used the reef resilience assessment to guide discussion of reef restoration, highlighting the importance of having such information available for adaptation planning. This study demonstrates how climate change impacts can be effectively incorporated into a management plan at the most granular level of planning and how a structured, formalized process can be as valuable as the resulting adaptation information.
Introduction

Natural resource managers and local communities need to make both near-term (years) and long-term (decades) decisions about how to manage environmental challenges. The effectiveness of these decisions increasingly depends on successful adaptation to climate change, which can affect management targets directly and through complex interactions with non-climate stressors [1–4]. Climate change adaptation has been receiving increasing attention in response to societal concerns regarding adverse impacts to species, ecosystems and human well-being. Much of the adaptation guidance developed so far has been generic and aimed at high-level policy rather than on formulating specific actions that are “climate-smart”. Structured approaches to adaptation planning that integrate existing methods for vulnerability assessment with design and evaluation of effective adaptation responses are needed [5,6]. This is because management plans and actions that are well designed for current climatic conditions may not perform as well under future conditions, such that the preferred course of action may differ when climate change is considered. All natural resource management actions should be assessed for their long-term effectiveness and made climate-smart by accounting for potential climate change effects in their design, implementation, and maintenance [5,6]. Practical efforts to operationalize climate-smart adaptation require consideration of climate change impacts throughout the natural resource management planning and implementation process. A key component is collaboration between communities, scientists, and managers in design of the climate-smart management actions.

The climate-smart planning cycle (Fig 1) [6] assists natural resource managers with systematically addressing the challenges associated with adapting to climate change (e.g., potentially large uncertainties around specific effects in a given location, the timing of climate change effects, and indirect effects) during each step of the planning process. Tools, guidance and numerous case studies exist for parts of the climate-smart management cycle, such as Step 2
vulnerability assessments [7–9] and resilience assessments [10–14]. However, there is less guidance on identifying and designing adaptation options for consideration (Step 4), or how to engage in more detailed climate-smart design of selected actions with experts in preparation for implementation (Step 6) [6]. These two steps are the focus of this study.

One available resource for identifying and designing climate-smart adaptation options in Steps 4 and 6 is the Adaptation Design Tool (ADT) [15]. The ADT lays out a stepwise process structured through worksheets by which natural resource decision makers can use information collected from local vulnerability and resilience assessments to link management actions to climate-smart adaptations. Through a series of targeted questions about the effects of climate change on stressors being managed and system responses, the ADT can help scientists, natural resource managers, and other stakeholders work together to apply the best available science to brainstorm and design effective actions. The objective is to empower managers to evaluate and select priority actions (Step 5, Fig 1) based on best-available information that proactively takes climate change into consideration, while also identifying and documenting information gaps, uncertainties, interactions among actions, and the sequences in which management actions should be executed [16].

In the ADT (Fig 2), Activity 1 helps managers and planners assess the effects of climate change on proposed management actions and adjust their designs to be more robust under changing conditions. Users list each management action under consideration and respond to questions about two categories of design considerations: 1) how climate change may affect stressors that are being managed by, or could affect, the action (worksheet 1A); and 2) what these indirect effects, along with any direct destructive effects of climate change on structural components, may mean for the effectiveness of the management action and how it could be modified to remain effective (worksheet 1B). Activity 2 (worksheet 2) then aids the brainstorming of new actions based on climate impacts and vulnerabilities that have not yet been addressed. The new actions are in turn passed through Activity 1 for analysis and design. The ultimate outputs of this structured process are descriptions of climate-smart management actions that can be evaluated (in Step 5 of the climate-smart planning cycle) instead of the actions that do not explicitly incorporate climate change, thus ensuring that climate change effects are considered in management decisions. While the ADT does not include analysis of cost-benefit ratios, political or social limitations, or other non-climate factors, insights on these issues are documented in the area for notes in Activity 1 and then considered during Step 5 evaluation and selection of priority actions. Once priority actions have been selected,
the ADT can be used again with subject matter experts to engage in more detailed implementa-
tion planning and design.

Although the ADT was developed and tested with the help of coral reef management practi-
tioners using example actions in specific locations [6,15], its implementation in an actual man-
agement planning context has not been formally explored. In this case study, we concentrated
on Activity 1 of the tool, examining existing brainstormed actions in a concrete management
situation. This proof-of-concept exercise aims to increase the tool’s utility by revealing strengths
and identifying areas for improvement during real-world implementation. It also demonstrates
how climate change can be integrated into environmental decision making, rather than handled
as a separate process. If climate change was not incorporated into the original management
plan, a good time to incorporate climate change is during management plan revision because,
during this time, activities are being reevaluated, reprioritized, and redesigned. As priority
actions are being considered for an updated plan, designing actions to be climate-smart from
the beginning may be easier and more economical than retrofitting adaptation into them later.

This study focused on a management plan revision for the Guánica Bay watershed and asso-
ciated coastal habitats in southwest Puerto Rico, with special focus on coral reef ecosystems.
The watershed and coastal area are relatively well studied with active management activities.
The current watershed management plan addresses a variety of threats (see Site Description
section in Methods), many of which are expected to be affected by climate change. Thus, this
was a good opportunity to incorporate climate change considerations in a planning cycle
using the ADT, covering both watershed and coral reef management activities. As part of this
process, we simultaneously explored the connection between adaptation planning and coral
reef resilience, using a recently developed Puerto Rico-wide climate change reef resilience
assessment [17] that we modified to focus on southwest Puerto Rico. Reef resilience assess-
ments use reef attributes that are indicators of sensitivity and adaptive capacity to estimate the
relative resilience of sites and present spatial information on surveyed reef locations [13]. Like
vulnerability assessments, resilience assessments inform management decisions, programs,
and activities, and are thus part of the climate-smart cycle and essential inputs to the ADT.
However, this is the first case where resilience assessment information and this tool have been
applied together in a structured process. In this paper, we describe how the ADT and resilience
assessment were used in an actual watershed management planning cycle, report on resulting
improvements to adaptation designs, and discuss lessons learned from the case study.

Methods

Site description

The Guánica Bay watershed (390 km²), located approximately 32 kilometers west of Ponce
and 160 kilometers southwest of San Juan, Puerto Rico, is dominated by evergreen forest (51% of
land cover) and grassland (26%), with the remaining watershed consisting of agricultural
land (e.g., cultivated land, pasture/hay, coffee), scrub/shrub habitat, developed land, wetland,
and bare land. Guánica Bay itself receives fresh water primarily from the mouth of the Río
Loco at the town of Guánica near the northern end of the bay. In the 1950s, five reservoirs and
two hydroelectric plants were built on the mountain ridges north of Guánica Bay to increase
and regulate potable water from the high elevation watersheds of the central mountain range
for use by the populations of coastal cities [18–20]. Canals and streams were also constructed
to divert water from just below the southernmost reservoir along the foothills to the west to
provide agricultural irrigation in the Lajas Valley. A long drainage channel was constructed
along the southern edge of the Lajas Valley to return the water eastward into the Río Loco near
its mouth [21–24] (Fig 3).
The Guánica Bay Watershed includes several existing conservation areas, including the Guánica Dry Forest (which is both a state forest and a United Nations International Biosphere Reserve), the Punta Ballenas Reserve (which is along the coast of the Guánica Forest and is managed as part of the Guánica Dry Forest), Cayo Aurora (commonly known as Gilligan’s Island also managed as part of the Guánica State Forest), and the Susúa State Forest (located between Yauco and Sabana Grande in the foothills of the Central Range). There is also the Lajas Valley Agricultural Reserve, which is managed for agricultural use. The Guánica Lagoon is also currently managed for agricultural use but has been identified by the Puerto Rico Department of Natural and Environmental Resources as a priority for wetland restoration.

Prior to the 20th century, the coral reefs of Puerto Rico and the wider Caribbean were dominated by branching corals, with structure and diversity enhanced by several species of mound-forming corals as well as sea urchins, large schools of game fish, and abundant sharks, turtles, and marine mammals [27]. Coral reefs still form extensive structures at the mouths of coastal embayments such as Guánica Bay and fringe many small islands along Puerto Rico’s southern coast [28–30]. However, these coral reefs have been significantly degraded over the past four decades by the cumulative effects of global and local stressors [2,30,31]. Local stressors (e.g.,
sediment, nutrient, and contaminant efflux from human activities in the adjacent watershed) have affected the reefs near Guánica and La Parguera (to the west), with dramatic reductions in living colonies of reef-building stony corals [32–34]. These losses have been further exacerbated by the impacts of rapid environmental changes and extreme events, such as hurricanes, altered precipitation patterns, droughts, ocean acidification and warm-water bleaching events [29,35].

In 2005, the U.S. National Oceanic and Atmospheric Administration’s (NOAA) Coral Reef Conservation Program contracted the Center for Watershed Protection (CWP) to develop a watershed management plan for the Guánica Bay watershed [21]. The plan identified the principle sources of pollution that threatened coral reef habitats in southwest Puerto Rico: upland erosion in coffee farms, reservoir sedimentation and transport, in-stream channel erosion, loss of the Guánica Lagoon, legacy contaminants, and inadequate sewage treatment. A series of management actions were identified to target these sources of pollution. In 2018, a follow-on report [36] summarized the activities that had taken place since the original management plan was released and provided a list of recommended actions for the future protection and restoration of the watershed and coral reefs. A revised management plan for the Guánica Bay watershed is currently being developed (R. Viquiera Rios, personal communication).

The U.S. Coral Reef Task Force (USCRTF) has identified terrigenous sediment as a major stressor to coral reefs at large and determined that its reduction is essential for maintaining coral reef resilience [37,38]. In southwest Puerto Rico, mass forest clearing, poor soils, and runoff associated with dirt roads, particularly in the mountain ridges, has resulted in increased sediment deposition and transport in the Guánica Bay watershed [39]. At the scale of individual farms, dirt roads have been found to be responsible for more than 90% of the annually generated sediment [40]. For this reason, management of sedimentation from dirt roads (e.g., water conveyance, diversion and flow reduction) is a priority and was examined in greater detail by an expert panel using the ADT (see Application of the Adaptation Design Tool section below).

NOAA’s National Marine Fishery Service (NMFS) released a recovery plan for elkhorn (Acropora palmata) and staghorn coral (A. cervicornis) in 2015 that included several adaptation measures, including increasing monitoring of disease and bleaching events, reducing local impacts of temperature stress (e.g., shading of reefs, pumping cooler waters onto reefs), researching the viability of land-based rearing and wild re-stocking of coral species, and testing approaches to culture resistant and/or resilient strains of corals (e.g., disease or biotoxin resistance, thermal or pH tolerance) [24]. For this reason, coral restoration, including land-based rearing (coral nursery) and wild re-stocking of species (outplanting), were examined in greater detail by an expert panel using the ADT (see Application of the Adaptation Design Tool section below).

Puerto Rico has experienced a variety of climate change effects which are expected to continue this century (S1 Table). Air and sea surface temperatures have increased throughout Puerto Rico, with sea surface temperatures warming faster in southern Puerto Rico than in northern Puerto Rico. Sea level has also risen, while precipitation trends are unclear so far.

### Application of the Adaptation Design Tool

Prior to applying the ADT, we compiled a list of potential management actions being considered for the updated Guánica watershed and reef management plan. These actions were drawn from the 2008 watershed management plan [21], summaries of multiple Guánica stakeholder workshops [22,23], and conversations with local managers about their priorities for the new plan (R. Viquiera Rios and P. Sturm, personal communication). This yielded a list of 80
potential actions, from which we selected using the following criteria: (1) adequate information and specificity about the proposed action and (2) at least some potential effect of climate change on the action. Examples of the types of actions that were excluded based on these criteria included developing rainwater collection systems (non-specific) and improved secondary school education about the importance of coral reefs (the education itself is not directly affected by climate change, though the content is). Using these criteria, we pared down the list to 14 actions (Table 1).

Next, we applied the ADT in two phases. The goal of the first phase was to develop basic climate-smart information for more robust evaluation and selection (Step 5 of the cycle) of priority actions through an initial screening process, corresponding to Step 4 of the climate-smart cycle. The goal of the second phase was to work with panels of subject matter experts to develop more detailed climate-smart designs for three of the management actions, corresponding to Step 6 of the climate-smart cycle; these would later be discussed with community members prior to implementation. Through these two applications, we were able to more comprehensively evaluate the tool in a real-world context.

Of the 14 actions in Table 1, actions 1 through 8 were related to watershed management, actions 9 and 10 to aquatic ecosystem protection, and actions 11 through 14 to aquatic ecosystem restoration. Actions 11 and 12 focused on the establishment of land-based fish and coral nurseries; although terrestrial, nurseries were identified as needing adaptation to climate change because of the threat to infrastructure posed by increasingly powerful storms. In the case of actions 13 and 14, there was insufficient information on coral reef resilience to guide decisions on fish releases and outplanting of corals during the first phase, so action 14 was only addressed during the second phase when resilience information became available (see Coral reef resilience assessment section below). Action 13 was never run through the tool.

| Action number | Action |
|---------------|--------|
| 1             | Plant cover crops in Guánica Valley farms |
| 2             | Plant riparian buffers along the Rio Loco where it passes through farms |
| 3             | Replace sun-grown coffee with shade-grown coffee |
| 4             | Hydroseed bare soils associated with roads and homes |
| 5             | Construct swales to treat urban stormwater (as a type of green infrastructure) |
| 6             | Restore Guánica Lagoon |
| 7             | Use water diversion structures and flow reduction practices (e.g., water bars, vetiver and rock check dams, culverts) to manage sediment from existing dirt hills and mountain roads |
| 8             | Construct wetlands for tertiary treatment at the Guánica WWTP |
| 9             | Protect seagrass meadows |
| 10            | Protect mangrove forests |
| 11            | Capture larval fish of target species and establish reef fish aquarium-based nurseries |
| 12            | Collect corals and establish aquarium-based coral nurseries |
| 13            | Release nursery-raised fish (from action 11) on reefs |
| 14            | Outplant nursery-raised corals on reefs (from action 12) around Gilligan’s Island to protect the coastline |

Once the resilience assessment was available, it was possible to address action 14 along with action 12 in the second phase because they were so closely related to each other. Actions 7, 12, and 14 were covered during the second phase. Action 13 was not addressed in either the first or second phases.

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One of us (DAG) completed the ADT Activity 1 worksheets for the actions 1 through 12 during the first phase, taking 3–4 hours on each action. Resources used included (1) the existing watershed management plan [21], (2) local reports and background resources on each action, (3) a Puerto Rico climate change vulnerability assessment (prepared by author PB using Puerto Rico Climate Change Council reports [41] and based on the Local Early Action Planning Tool [42] (S1 Table), and (4) focused discussions with local experts for some information on potential actions (e.g., the design status of the wastewater treatment plant’s tertiary filtration wetland and the history of Guánica Lagoon). The tool was applied iteratively to the actions because responses to one action sometimes influenced or informed others. For example, several of the actions related to watershed management were intended to manage the same stressors, so information gathered for one action could sometimes be applied to others. For some of these watershed actions (e.g., water diversion techniques to reduce road erosion), the sole stressor being managed was sediment, while nutrients and toxic chemicals were additional stressors to be managed by other actions (e.g., restoring Guánica Lagoon).

To synthesize the output of the two ADT worksheets, we developed five-column summaries that consolidated outputs for each action into a row that could fit on a single page, rather than on the two pages generally required for the full ADT worksheets. The Guánica Bay watershed managers felt that these five-column summaries would be a preferable format for decision makers and community members who may prefer a more concise synopsis. They suggested that the five-column summaries be included in the bodies of reports in general and the full ADT output be placed in report appendices.

For the second phase, we selected two of the initial 12 management actions (Table 1) that were being implemented or that local managers were highly confident would be implemented: dirt road management in mountainous farms (action 7), and establishment of coral nurseries (action 12). With the reef resilience assessment now available (see next section), it also was possible to examine coral outplanting (action 14) as a closely related reef restoration activity supported by coral nurseries. Both dirt road management and coral reef restoration are longstanding priorities for the Guánica area because mountainous roads are recognized as major contributors to stream and coastal sediment loads [40,43–45] and there is a strong emphasis in the community to improve the health and recovery of reef communities.

The approach for this phase involved working with local experts to elicit the more detailed level of adaptation information needed for implementation planning. Two panels of six experts (one for dirt road management and one for coral reef restoration) were consulted (S2 Table). Experts were selected on the basis of having complementary skills and knowledge of southwest Puerto Rico, as well as combining local actors with scientists. They were fully aware of the existing conservation areas and existing management efforts.

We used the following structure for the discussions. In an initial 90-minute web meeting, we introduced the ADT and the case study. We also provided the results from the first phase on which the second phase would be built, plus the recent preliminary results of the resilience assessment of Puerto Rico’s coral reefs (see next section). We followed the introductory call with two 90-minute web discussions a few weeks apart, during which the authors facilitated discussion of the actions following the structure of the ADT worksheets. Our goal was to elicit as much information on the actions from the experts as possible, rather than achieve consensus among experts. The authors updated the ADT worksheets in real time using the conference system’s screenshare, enabling the experts to share their respective opinions on and reactions to proposed text in the ADT worksheets. During the two web discussions for each panel, we added increasingly more detail to the information from the first phase, with some revising of the focus of the first phase actions. Experts reviewed successive drafts of the worksheets and their comments were collated within the ADT output.
Coral reef resilience assessment

Concurrently with the ADT case study, a desktop resilience assessment of Puerto Rico’s coral reefs was conducted [17] using data from the most recently completed NOAA National Coral Reef Monitoring Program (NCRMP) survey of Puerto Rico. Based on the methods of [46], it involved selecting seven expert-vetted indicators of coral reef resilience to ocean warming associated with climate change (percent live coral cover, coral diversity, percent algae cover, incidence of coral disease, thermal tolerance of hard corals, biomass of herbivorous fish, and benthic rugosity), then rescaling them so that all indicator sites (n = 103) were assigned fractional values between 0 and 1, with 1 being the site most resilient to thermal stress. The indicators at each site were averaged to calculate a relative resilience score for each site, and sites were ranked by resilience (S1 Fig). Each site was also assigned a relative score for the stressors of fishing pressure [47] and land-based sources of pollution, with the values scaled the same way as the resilience indicators. Land-based pollution was modeled using OpenNSPECT [48] for watersheds and a simple coastal dispersion model to estimate relative sediment loads at reef sites. Indicator values, stressor values, and the resilience score were all relative to the surveyed sites, not based on any absolute scale. Using this assessment, reefs most suitable for specific management actions (e.g. reef restoration or mitigation of land-based sources of pollution) can be identified by defining criteria for those actions based on the individual indicators, stressors, and/or resilience scores (called "management action queries"). For more information, refer to [17].

We presented a preliminary version of this Puerto Rico-wide assessment to both expert panels as potential input for their use of the ADT, with the intent of assessing how the resilience assessment helped inform the design of the management actions that were considered in detail (S1 Fig). Based on feedback from the coral reef restoration panel when they discussed the two coral restoration-related actions, we localized the Puerto Rico-wide assessment to the area surrounding Guánica Bay. For this, we used the same methods as [17], but included just the 15 NCRMP survey sites in southwest Puerto Rico that had the full suite of collected survey data, and rather than rescaling indicators and resilience to the "best" value for all of Puerto Rico, we rescaled among the 15 local sites. This approach allowed the Guánica sites to be considered relative to each other (rather than sites throughout Puerto Rico) region, making the approach more relevant for Guánica-area planning (S1 File and https://github.com/dagibbs22/Puerto_Rico_Resil_assmnt).

Results

Adaptation Design Tool

The main output of this case study is the adaptation information from the first and second phases of using the ADT, as well as the comparison between them. Tables 2 and 3 present combined and condensed versions of both the first phase (black text) and second phase (blue text) results for the selected actions. Table 2 summarizes the climate change effects on stressors (worksheet 1A), while Table 3 summarizes the climate change effects on the management action (worksheet 1B). The information from both worksheets culminates in Column B7, which is the climate-smart design (highlighted in Table 3). The full-length versions of the first phase results for all 12 actions are available in S3 Table, the results for the second phase are in S4 Table, and the five-column summaries for the 12 actions evaluated in the first phase are in S5 Table. Uncertainties in and limitations of existing data are found in the “Notes” columns of all tables.

First phase results. Because the eight watershed-focused actions addressed similar stressors, the tool outputs were similar for the effects of climate change on stressors (worksheet 1A);
Table 2. Worksheet 1A from the Adaptation Design Tool.

| A1 | A2 | A3 | A4 | A5 | A6 | A7 |
|----|----|----|----|----|----|----|
| Management action number | Existing management action (e.g., water bars, vetiver and rock check dams, culverts) to manage sediment from coffee plantation dirt roads. | Target stressor(s): Terrestrial sediment | Climate change effects on stressor(s): direction, magnitude, mechanism, uncertainty. | Impact of climate change effects: | Implications for effectiveness metrics and how to measure them | Notes |
| 7 | Use flow diversion structures and flow reduction practices (e.g., water bars, vetiver and rock check dams, culverts) to manage sediment from coffee plantation dirt roads. | Terrestrial sediment | • Storms may become more intense, leading to precipitation events that may erode dirt roads faster, particularly on steep slopes. The percent increases in erosion and runoff will likely be greater than the percent increase in precipitation. High magnitude, low uncertainty. | • The rainfall threshold for erosion of dirt roads appears to be around 0.1 cm, with dirt released within 1–2 minutes (Ramos-Scharron and Thomaz 2016). This threshold may be reached during a higher proportion of storms. High magnitude, low uncertainty. | • Storms of sufficient intensity to release levels of sediment exceeding reefs’ tolerance may occur more frequently. High magnitude, medium uncertainty. | • Increased resuspension of sediment in coastal waters in between storms due to stronger waves. High magnitude, medium uncertainty. | • More intermittent storms could cause larger loads per event because drier soil will erode more easily. On the other hand, hard pans could form on roads during dry periods, causing initial rain to run off without sediment. Medium magnitude, medium uncertainty. | • There may be shifts in eroded sediment particle sizes due to larger storms, potentially including more clay particles which stay suspended in water longer and chronically expose reefs to sediment. Or, larger particles may be increased. Low magnitude, high uncertainty. | • Increasingly violent storms are already occurring. | • Storm intensity will likely continue to increase over the coming decades. | • Prolonged dry periods are already occurring. | Effectiveness metrics: Targeted percent reduction of sediment loads originating from dirt roads in coffee plantations. Reduction in number of roads that need rebuilding or significant management after storms. Amount of sediment resuspension. Frequency of roads requiring reggrading. Effective lifespan of conveyance structures. | • How much are 2-, 5-, 10-, and 25-year storms expected to change by 2050? | • What is the total sediment runoff reduction target for reefs? | • How much of a reduction in runoff from use of road maintenance practices is necessary to reduce runoff to levels that reefs can tolerate in conjunction with other management measures? | • How do road maintenance measures interact with other mountain erosion measures (e.g., shade-grown coffee) to reduce sediment? | • Roads really need to be monitored during storms to detect locations and timing of most severe sediment runoff. | • No management measures will protect dirt roads from large storms. There is a limit on how large a storm event dirt roads can be designed to handle. That limit may be reached more often under climate change. | • Residence times of sediment particles around reefs outside Guánica Bay could determine the relative contribution to coral exposure of sediment resuspension vs. new sediment. | • How will climate change affect the size of sediment particles mobilized? | • Information on changes in runoff coefficients are necessary. | • Pattern of sediment pulses on coffee plantation dirt roads under climate change is very important and needs exploration. | • How will climate change affect the size of sediment particles mobilized? | • Information on changes in runoff coefficients are necessary. | • Pattern of sediment pulses on coffee plantation dirt roads under climate change is very important and needs exploration. | |

**NOTE:** Some flow diversion practices, and road stabilization measures are included with water conveyance practices because implementing them together increases their effectiveness.
| Existing management action number | A1 | A2 | A3 | A4 | A5 | A6 | A7 |
|-----------------------------------|----|----|----|----|----|----|----|
| **Management action**             | Existing management action number | Target stressor(s) | Climate change effects | Implications for effectiveness metrics and how to measure them | Management | Timing of climate action number: direction, magnitude, change effects | Implications for how to measure effectiveness | Notes |
| 12 First and second phases: Collect corals and establish aquarium-based coral nurseries. | | | | | | | |
| 14 Second phase only: Ship cable | | | | | | | |

This covers climate change’s effects on stressors. First phase text is in normal text. Material added during the second phase is in italics. Text is condensed and edited for this table, and actions are numbered as in Table 1. The full worksheets from the first and second phases, including column heading descriptions, are in S3 and S4 Tables, respectively. For condensed summaries of the initial phase, refer to S5 Table.
### Table 3. Worksheet 1B from the Adaptation Design Tool

| B1 | B2 | B3 | B4 | B5 | B6 | B7 | B8 |
|----|----|----|----|----|----|----|----|
| **Management action number** | **Existing management action** | **Changes in effectiveness of management action due to climate impacts on target stressor** | **Changes in effectiveness of management action due to climate impacts on management action** | **Time frame or constraint for using the action and implementation (e.g., urgency, longer or shorter term)** | **What changes are needed to adapt the action (place, time, engineering design)** | **Climate-Smart Management Action** | **Notes** |
| 7 | Use flow diversion structures and flow reduction practices (e.g., water bars, water conveyance, culverts) to manage sediment from coffee plantation dirt roads | Use water conveyance practices (e.g., water bars, culverts) to manage sediment from coffee plantation dirt roads | Use water conveyance practices (e.g., water bars, culverts) to manage sediment from coffee plantation dirt roads | Use water conveyance practices (e.g., water bars, culverts) to manage sediment from coffee plantation dirt roads | Use water conveyance practices (e.g., water bars, culverts) to manage sediment from coffee plantation dirt roads | Use water conveyance practices (e.g., water bars, culverts) to manage sediment from coffee plantation dirt roads | Use water conveyance practices (e.g., water bars, culverts) to manage sediment from coffee plantation dirt roads |

**NOTE:** Some flow diversion practices, and road stabilization measures are included with water conveyance practices because implementing them together increases their effectiveness.

- Water diversion structures may not be able to divert all water off roads, producing downhill road erosion.
- Flow and erosion reduction practices may not be able to slow down water sufficiently to prevent further erosion and trap sediment.
- Vertical dirt walls along roads may be more likely to collapse under increased precipitation.
- Sediment trapping structures may not be able to handle the resulting increased loads.
- Intertensions between stones in check dams or stones swales may become clogged with sediment more quickly from larger or more frequent storms. This will reduce the cross-sectional area available for water to flow through and the area available for conveying water.
- Larger storms may wash out existing water conveyance structures (e.g., water bars, the stones comprising check dams, or culverts).
- Larger storms may wash out existing flow and erosion reduction practices (e.g., water and rock check dams).
- Larger storms may overflow culverts or wash them out entirely, leading to stream channels being washed out.
- Road regrading may be complicated by increased risk of raised road banks slipping onto road during or after regrading. Slippage is already happening but could happen more under some precipitation scenarios.
- Dodging may compromise the effectiveness of vegetative solutions. Vetiver probably does fine in droughts, native plants (investigated by the US Fish and Wildlife Service) may not do as well.
- Work completed less than one week to one month before a large storm may be undone, i.e., a completed project can take a month to establish. Depending on the pattern of large storms, this could increase, decrease, or simply shift the window for a splitting projects.

This suite of actions can be implemented immediately. They have the potential to quickly affect sediment loads.

- Compact dirt roads that have been built to handle storms of a specified size (e.g., predicted 5-year storm).
- Reduce the cross section of the road to some standard minimum size when constructing the project. Maintenance will have to be more frequent.
- Use vegetation that are withstand both dry periods and stronger flows.

Minimize sediment from existing dirt mountain roads by building water diversion structures more frequently along roads, sloping roads more heavily to promote faster drainage, and increasing barriers on upstream sides of culverts and flow diffusers on downstream sides of culverts. Culvert size should be increased to a standard minimum size in preparation for consistently larger flows. Locations requiring flow control may change due to altered precipitation patterns. Check integrity and repair diversion structures after larger storms; remove sediment from sediment traps after large storms. Compact roads with surfaces made of small rocks and granular material to stabilize road surfaces. Paved roads that have already been repeatedly washed out with other mitigation techniques are not possible.

- These structures and practices are considered together (as a suite of actions) because they must be implemented in combination in order to be effective.
- To what extent can dirt roads traversing hills and mountains be unpaved?
- Which roads are most important to work on?
- This is a feedback loop: the worse the road water management actions perform, the more erosion there will be, and the worse they will perform.
- When is paving dirt roads an option? Are they generally too ephemeral to justify paving?
- Dirt roads can be stable for a long time if managed properly.
- However, once they begin to degrade, they can fail apart very quickly. They have a failure threshold and maintenance keeps them from reaching that threshold.
- Road maintenance is not generally funded in grants, yet it is essential for long-term performance of structures. Thus, maintenance falls to the property owners (farmers). Pretty much all maintenance will be affected by climate change.
- No-cost extensions would increase flexibility in completing work under greater weather uncertainty.
- Climate change may reduce the lifespan of projects. Standard lifespan now is 25 years.
- Project locations are based on which farmers are willing to collaborate, as well as slope, potential flooding rate, and connectivity to water bodies, and traffic load. To what extent will climate change affect this?
- Supply farmers with policies of the right size so they are not just using whatever they have handy.
| B1 | B2 | B3 | B4 | B5 | B6 | B7 | B8 |
|----|----|----|----|----|----|----|----|
| Management action number | Existing management action | Changes in effectiveness of management action due to climate impacts on target species | Changes in effectiveness of management action due to climate impacts on management action | Timeframe or constraint for using the action and implementation (e.g., urgency, longer or shorter term) | What changes are needed to adapt the action (place, time, and engineering design) | Climate-Smart Management Action | Notes |
| 12 | First and second phases Collect corals and establish aquarium-based coral nurseries. | Fewer fragments of coral colonies (especially of certain species) will be available for collection after storms due to reduced coral cover. | The sooner this is done, the better | | | | |
| 14 | Second phase only Outplant strains on reefs around Gilligan’s Island to protect the coastline. | • Fewer fragments of coral colonies (especially of certain species) will be available for collection after storms due to reduced coral cover. | • The sooner this is done, the better | | | | |

First and second phases: • Fewer fragments of coral colonies (especially of certain species) will be available for collection after storms due to reduced coral cover. • On the other hand, opportunities for collecting new nursery stock may increase since stock will only be collected after storms (and at construction sites). • Available coral colonies may be more resistant to higher temperatures and existing diseases (through natural selection). • Propagating coral genotypes without regard to their resilience to climate stressors will reduce action effectiveness because individuals in the nursery will have the same tolerance of climate change conditions as wild corals. • Outplanting sites selected because they are currently in deep water may be in too deep water in the future due to sea level rise. • Turbidity or coastal erosion plans/keels may change in such a way that outplanting locations selected to avoid hotspots or plumes may no longer do so. • Outplanting locations not currently exposed to legacy sediment contaminants could become exposed from increased resuspension or runoff. • Decreased growth and reproduction rates of outplants may occur from ocean acidification. • Abiotic stress effects may reduce the genetic diversity of outplanted colonies below the intended level. • Climate change may physically damage or destroy nurseries through more powerful storms. • Nurseries must be prepared for power failures, loss of clean water, and other emergences. • Increased physical destruction of outplanted colonies by storms, especially for more delicate species. • Increased death of outplanted colonies from bleaching and diseases. • It will be 2-4 years before corals can be outplanted, so this is a medium-term action. It should probably be started immediately, to reduce long-term coral losses. On the other hand, if coastal water quality does not improve, outplanted corals may not survive. • The sooner this is done, the better • Outplanting will have to be a sustained, long-term effort. • Preferentially grow coral genotypes that are disease-, heat-, and sediment-tolerant/resistant. In general, grow genotypes that will be able to survive projected future conditions. As conditions change further on reefs (e.g., due to increased coral mortality from bleaching and disease, collect more colony fragments due to higher mortality of outplanted colonies. • The similarity between the water supplying the rearing aquarium and the place(s) where the corals will be outplanted also needs to be considered. Pessimistically, these should be similar in temperature, pH, and chemical composition. • Focus on outplanting coral strains with a variety of types of tolerance to climate-change effects, including being bred to tolerate multiple stressors. • May need to focus on outplanting at shallower sites due to sea level rise, although that must be balanced against the impacts of larger storm and shallower runoff shallower depths (to grow). • If outplanting in deep water, focus on using species with a broad depth range to account for sea level rise. • Outplant some aquaculture-bred colonies with the hope that they will be naturally fragmented and propagate themselves, and outplant others to build-up nurseries where they may be desired. • Cement may need to be used more often to attach colonies to substrate due to increased risk from storms. • Time outplanting to avoid periods of enhanced runoff and land-based pollution. • Factor changes in sediment phase location and direction into outplanting site selection. • Monitoring may need to be extended longer after outplanting to include observation of how outplants handle extreme events and to what extent they reproduce. Develop multiple species aquarium-based coral nurseries which can produce a continuous supply of coral colonies through repeated fragmentation. Species that can survive the current temperature, pH, sediment and nutrient regime should probably be used until water quality is restored. This may involve collecting new colony fragments that have survived widespread bleaching or disease or survived large sedimentation events. Coral strains should be heat-tolerant. To reduce risk of bleaching and some some resistance to the relevant coral diseases that are associated with higher temperature. They should also be effective at removing deposited sediment and maintaining growth in lower pH waters. Water used in the aquaria should be from the general area where the corals will be outplanted. Outplant colonies that are heat-, disease-, salinity-, and low pH-tolerant. Species mix should be optimal for robust coral defense. Outplanting during high sediment or precipitation periods should be avoided to reduce initial stress exposure to outplants. Balance outplanting locations between deeper and shallower sites, accounting for how water level rise may make deeper sites uninhabitable in the future and how shallower sites may be more heavily exposed to land-based pollution. Colonies in shallower sites may need to be attached to reefs using cement more frequently due to larger storms. Colonies in deeper sites should have a broad range of depth tolerances. In either case, they may be desirable to locate breeding colonies where they will be naturally fragmented. Site selection may also be affected by shifting plumes of land-based pollutants from the bay. Because survival through extreme events is an important part of the nursery program, colonies should be monitored through extreme events such as high temperature events and large storms. Monitoring periods for outplants may need to be extended to do this. • Preferentially grow corals that are disease-, heat-, and sediment-tolerant/resistant. In what order should these traits be prioritized or how should they be balanced? Many organizations are working on creating hybrids that combine these traits. • The nurseries must be large to have an ecologically meaningful amount of coral in them. How much is that? • Can bathymetry be used to predict where storms will be most or least damaging? • Refit the west of Guánica Bay receive more sedimentation than those to the east. Reef evidence assessment shows some potentially suitable sites to the east of the bay. • Most people are attaching colonies individually at this point. What can be done to successfully attach multiple colonies simultaneously? • Need to focus on other actions that are important for improving the outplanting environment. |
however, the outputs diverged for the effects of climate change on management actions and how to adapt (worksheet 1B) (S3 Table). For the most part, the watershed-focused management actions (e.g., management action 7 in Tables 2 and 3) centered on adapting to changes in precipitation patterns due to climate change and being able to mitigate impacts from both drought and flood conditions. Indeed, most of the actions had notes about similar information gaps in precipitation projections, suggesting that the same research could be used to fill knowledge gaps for many of the actions. The actions also had similar monitoring concerns, such as capturing action effectiveness under the full range of stressor loads during storms of increased intensity. The importance of adaptive monitoring to measure the effectiveness of actions was highlighted as a priority.

One key piece of information that was consistently missing from the watershed management actions was what the target land-based pollution reduction should be for a given action at a specific site, as well as for all actions in aggregate. Part of the uncertainty was due to a lack of information on the historical and projected effects of land-based pollution on coastal ecosystems, including coral reefs. Potential effects on coral reefs, and uncertainties about those effects, were included in the watershed actions insofar as climate change affects how runoff is distributed in the ocean (plume extent, frequency, particle size, etc.) and how often reefs' tolerance thresholds for land-based pollution might be exceeded.

The final outputs (descriptions of climate smart versions of the actions) included not just changes to actions' physical sizes (i.e., climate change adaptation does not simply require making the action physically bigger), but also included changes to siting and construction methodologies or materials. Additionally, many of the climate smart actions specifically identified adaptations that might be needed for maintaining the action, not just adaptations in the construction. For example, hydrosedding bare soils will require more follow-up to sustain soil coverage (S3 Table). Because the exact location of where most of these actions would be implemented was not specified in the sources they were extracted from, the adaptations were based on general principles for prioritizing locations (e.g., soil types or slopes). Finally, some of the climate-smart actions had to address the possibility of a wider range of levels due to fewer or stronger storms (e.g., restoration of Guánica Lagoon). This was accomplished by including recommendations for vegetation that were both drought-resistant and well-rooted.

The first-phase focus for coastal actions (seagrass and mangrove protection, creating larval fish nurseries, creating coral nurseries) was altering the location of interest (habitat protection) and target species and traits (fish and coral nurseries). As with watershed actions, the two habitat protection actions and the two nursery actions were more similar in worksheet 1A (effects of climate change on stressors) than in worksheet 1B (effects of climate change on management actions and how to adapt). For targeting the action location, emphasis was placed on maximizing intra- and inter-habitat connectivity under climate change, filtering/trapping land-based pollution, and identifying habitats that could serve as climate refugia while accounting for existing conservation areas (S3 Table). Because sea level rise is a threat to both mangroves and seagrasses, it was considered a protection criterion for both resources. For the nurseries (actions 11 and 12 in Table 1), the focus was on how climate change could affect continuous operation of the nurseries during extreme events, the acquisition of fish and corals for the nurseries, and what species and traits to genetically select for propagation. Terrestrial nurseries are expected to be threatened, and their operations interrupted, more often by large storms and hurricanes, and therefore require additional emergency abilities (e.g., how to handle lack of clean water and prolonged power failures). Aside from resilience of the nursery facilities themselves, the water in the aquaria should be as similar as possible to the water at the outplanting sites (e.g., temperature, pH, sediment loads). In general, adaptive traits for corals and fish were theorized to include lower sensitivity to high temperatures, tolerance of more
acidic water, and disease resistance (for corals). Finally, a key precondition for these two actions was noted: in situ water characteristics and habitat must be of sufficient quality to make undertaking these worthwhile. If local stressors are not addressed and there is no suitable habitat into which to release nursery-raised fish or corals, the effort put into the nurseries may be better directed elsewhere.

**Second phase results.** During the second phase analysis by the expert panel, the focus of the actions was refined and shifted based on new considerations (Tables 2 and 3, S4 Table). For dirt road management in coffee farms, the experts wanted to focus on a narrower set of actions, so they shifted from flow diversion and flow reduction practices to water conveyance practices. However, they noted that water conveyance, flow diversion, and flow reduction practices are often implemented concurrently, so they ultimately included a broader suite of interventions in their discussion. In order to show development of more location-specific detail using an example, the experts discussed a farm known to have owners who were interested in erosion control and would be favorably inclined to make their plans more focused on adaptation. While using that farm as a specific case, the experts also generalized their observation to other locations with similar conditions during the discussions.

After narrowing of the types of water management actions under consideration, the road management expert panel generally agreed with the first phase information but added new details as well as important research needs captured in the Notes (Tables 2 and 3). For example, they proposed a few situations where climate change could affect the system in divergent ways. One such situation was whether a hard crust of dirt (hard pan) might form during dry periods, thereby reducing sediment loss during some ensuing storms. They also discussed changes in sediment particle sizes due to changing precipitation patterns (including resuspension on coral reefs), although they concluded there was not enough information for them to make any predictions about how sediment particle sizes would change. They also emphasized the potential reduction in project lifespans under climate change due to more rapid wear and greater chances of destruction, and that those revised lifespans must be built into project planning and budgets. Similarly, they noted that maintenance needs for projects would likely increase and would need to be accounted for early in project development. Just as in the first phase, they were not able to assign a target percent reduction for sediment loads from dirt roads in coffee farms because too little is known about the system. Finally, prompted by their experience during Hurricane Maria (2017), they acknowledged that adaptation to storms above some intensity is infeasible, although they did not identify where that threshold would occur.

For the coral reef restoration action, the experts discussed both nursery creation (included in the first phase) and outplanting steps of coral restoration (not included in the first phase) because they are so closely related and because outplanting was of particular interest to coral reef managers in Guánica. The outplanting action had not been considered in the first phase because it was judged that a resilience assessment was needed for effective site selection before designing an outplanting scheme. However, with the help of the draft resilience assessment that was completed between the first phase and the second phase (see resilience assessment results below), the experts were able to choose an outplanting location and objective for restoration using the resilience assessment’s suggestion of suitable sites for targeting interventions. Hence, they expanded their focus to examine outplanting of corals (action 14 in Table 1) around Gilligan’s Island (Cayo Auroro) with the objective of protecting the coastline from erosion. Gilligan’s Island, located a mile off the coast of Guánica, is part of the Biosphere Reserve of Guánica, and is managed by the Puerto Rico Department of Natural Resources. Although there are no reef resilience assessment sites at Gilligan’s Island, it is relatively near and shares characteristics with one of the survey sites that was identified as suitable for restoration.
For the coral reef restoration second phase, the experts emphasized ocean acidification and legacy sediments more heavily than was done during the first phase analysis. They also noted that site conditions are changing directionally (lower pH, higher temperature) and that nursery conditions need to reflect those ongoing shifts. Their proposed success metrics included maintaining genetic diversity of corals, benefits throughout the full life cycle of corals, the community consequences of restoration (e.g., functional roles served by outplants for other taxa), and reductions in storm surge and coastal damage. The experts were more specific about the traits that corals will need under climate change, including heat tolerance, sediment removal, and maintaining growth in low pH water, and how there might be tradeoffs among these. They proposed conducting long-term monitoring that would capture effects from the rare, large storm or bleaching event and would be highly responsive to such events. In terms of the siting for outplanting, they noted the need to account for potential shifts in locations and extents of sediment plumes under climate change, although they did not have specific predictions or recommendations for the study area. A final theme was the depth for out-planting corals, recognizing tradeoffs between risks from increased storms and temperature at shallow depths versus limits of each species depth range.

Coral reef resilience assessment

The road design expert panel acknowledged that watershed management activities like road design affect coral reefs, but the group was primarily interested in practices to reduce overall erosion from dirt roads rather than how road specifications were related to reef resilience; thus they did not see a map of reef resilience as informative for road design. In contrast, the coral restoration expert panel did use the coral reef resilience assessment to identify areas of high and low resilience and areas that might be good restoration opportunities based on pre-determined criteria for selecting coral restoration sites (management query, per [17]). However, they noted that the Puerto Rico-wide coral reef resilience assessment was too large-scale for reef actions focused on Guánica, prompting us to localize the resilience assessment to just the Guánica region (Fig 4). This approach allowed the Guánica sites to be considered relative to each other (rather than relative to all sites throughout Puerto Rico), making the approach more relevant for Guánica-area planning. One of the two sites that fit the criteria for coral restoration was just to the east of the mouth of Guánica Bay (number 3 in Fig 4), not far from Gilligan’s Island, the site that the restoration panel considered for the second phase. It likely would not receive much land-based pollution from the Guánica watershed because the major ocean currents in this area flow to the west. The next steps after identifying this specific survey site as having high potential for restoration based on resilience would be to discuss with stakeholders and confirm benthic habitat suitability through field investigations.

Discussion

This is the first application of the ADT during a management planning exercise. To that end, selecting potential management actions, running them through the tool as a first phase, and determining how best to summarize them (e.g., five-column summaries, S5 Table) for use in evaluation and prioritization was novel. Likewise, this is the first application of the ADT to explicitly produce the detailed level of information needed to carry out the implementation step of the climate-smart planning cycle (Step 6 of the climate-smart planning cycle, Fig 1). This dual use of the ADT allowed for a comparison between the pre-climate-smart management actions, first phase management actions made partially climate-smart by a single non-expert, and second phase actions made more thoroughly climate-smart by groups of experts. The generalities of the first phase outputs lend themselves to more regionally applicable
guidance on making management actions climate smart that can be further customized and elaborated upon for specific locations; the outputs could serve as reference material for related actions in similar situations. We also identified the extent to which the ADT can contribute to planning for implementing actions; experts can provide significant details towards implementation while using the tool, assuming the action is sufficiently specific in location and objective. These outputs need to be further discussed with local stakeholders if they have not already been included during the use of the ADT to make sure stakeholder knowledge is reflected and buy-in is achieved. The integration of stakeholder knowledge with expert use of the ADT during project implementation is an opportunity for further research.

The two expert panels found the ADT worksheets useful for systematically evaluating the potential impacts of climate change on intended management actions and for working through adaptation strategies. During the two 90-minute discussion sessions for each panel, the experts added considerable detail to the first phase of the actions and moved toward plans for action implementation for specific locations in the Guánica region. The experts provided in-depth knowledge on the limits of management actions (e.g., when it is just not possible to have a large enough culvert to mitigate more intense/frequent storm runoff) and the timing needed for actions to be effective.

We also explored how spatial resilience assessments could be used in conjunction with the ADT. The climate-smart cycle indicates that information on resilience along with vulnerability are inputs to Step 4 of the climate-smart cycle (identifying adaptation options); however, direct
use of a resilience assessment in identifying adaptation options had not previously been tested. While spatial patterns of coral reef resilience were not deemed relevant to road design, the coral reef restoration expert group used the island-wide resilience assessment to discuss outplanting design, providing insights into the connection between spatial resilience data and the ADT. There are two general ways in which resilience data can inform use of the ADT: site selection and identifying and designing climate-smart management actions.

First, maps of relative resilience can be used to select priority sites for management activities, whether the objective is to maintain resilient sites or to improve less resilient sites. For example, the survey site just east of the mouth of Guánica Bay has the third-highest resilience based upon multiple contributing factors. The prevailing ocean currents are driven by trade winds and flow westward, transporting effluent from Guánica westward towards La Parguera [31,49]. The Guánica State Forest (Bosque Estatal de Guánica) and UNESCO Biosphere Reserve extend to the east of Guánica Bay. The Guánica Dry Forest has been a protected ecosystem with minimal anthropogenic land use (United Nations Biosphere Reserve since 1981 and a commonwealth forest since 1917) [50]. The Guánica State Forest also manages Punta Ballenas Reserve, which contains mangrove forest, submerged aquatic vegetation and coral reefs, and the Cayos de Caña Gorda, a group of three uninhabited, mangrove-covered cays. The natural, native terrestrial landscape, protected mangroves, seagrass and coral reefs, combined with the westward ocean currents contribute to the high resilience of these coastal systems. Aware of these conditions contributing to the relatively high resilience of the site, the experts chose one of the nearby three cays (Gilligan’s Island) as the focus for reef restoration (Fig 4).

Second, in terms of identifying and designing actions, the management queries of the resilience assessment help managers identify which actions can maintain or improve resilience at a given site. The ADT can then be used to design those actions so that they are successful under current and future conditions. For example, the Gilligan’s Island site is indicated for coral restoration, so restoration interventions in this location would then be put through the ADT. Information on the relative resilience of candidate sites—and what is conferring that resilience—also can be essential for aiding climate-smart management action design. For example, ADT worksheets for activities involving acquiring resilient corals (e.g., collecting corals for nurseries and outplanting, or identifying potential refugia for protection) ask for what is known about the location of resilient corals and/or the condition of the site with regard to key factors that support resilience (such as healthy herbivore populations). This explains why, in this study, the resilience maps were not as helpful for experts whose main objective was reducing sediment loads throughout a watershed that had already been selected (e.g., Guánica Bay). However, the resilience maps were of greater interest to the expert group focused on setting up nurseries and outplanting corals; for them, the resilience maps provided some clues as to where resilient corals might reside and be collected for propagation, and what sites are already resilient and ready to receive outplants.

A different extent of resilience assessment might be needed when considering how to reduce sediment throughout the watershed. In natural systems, multiple processes operate simultaneously at numerous spatial and temporal scales [51–53], and patterns or relationships observed at one scale may be invisible when examined at another scale [54]. Climate change adaptation requires consideration of the broader landscape context [5]. In this case, consideration of the number and location of dirt road restorations required to protect/increase coral reef resilience could be a follow-on study. Additional information at the watershed scale would be required, including locations of all dirt roads, identification of which streams would be receiving the stormwater runoff from the dirt roads and how that would affect the streams themselves, and how much water could be diverted from the dirt roads under various rainfall
scenarios. This could be added to the resilience assessment as a best practice for those facilitating the process. The alignment of scale and extent between the resilience assessment and ADT is one difficulty that future researchers and managers should consider early in the management cycle.

Finally, it is important to note that use of resilience assessments and the ADT is iterative. As resilience assessments are updated, it can signal to managers when site selection and selected actions need to be re-evaluated. Likewise, as new actions or locations are evaluated and selected, a new resilience assessment may be needed, including updated and/or new types of data. The management action designs crafted with the tool are only as good as the information available to managers (e.g. vulnerability and resilience assessments), but the ADT can be used iteratively as new information becomes available.

In conclusion, the use of the Adaptation Design Tool in conjunction with a coral reef resilience assessment is a proof-of-concept consistent with adaptive management, a key component of which is to proceed with the best information available and work iteratively. We have demonstrated how a spatial resilience assessment can (1) contribute to identifying sites at which to prioritize adaptation actions and (2) help with designing actions. The ADT leverages existing information to more effectively protect and restore valuable natural systems such as watersheds and coral reefs, and combines that information to provide insights into adaptation. As the expert panels noted, the value of convening an interdisciplinary group of experts to tackle a complex problem using a logical decision process is as valuable as the resulting information itself. This process of combining multidisciplinary information on resilience, vulnerability and their implications for adaptation design supports robust decision making within the context of the uncertainties associated with climate change, by using the best available information while allowing the flexibility for continued improvements in the future.

Supporting information

S1 Fig. Puerto Rico-wide coral reef resilience assessment. The 103 sites are from the NOAA National Coral Reef Monitoring Program 2014 survey. Sites are shown by resilience rank (1 is most resilient, 103 is least resilient) and resilience quartile. Reprinted from [17] under a CC BY license, with permission from PLoS One, original copyright 2019.

S1 Table. Vulnerability assessment for Guánica Bay watershed and associated coral reefs (prepared by author PB for expert consultations described in [15]).

S2 Table. The experts who participated in the second phase panels. Institutional affiliations are those at time of panel meetings (summer 2017).

S3 Table. First phase output for Adaptation Design Tool Worksheets 1A (effect of climate change on stressors) and 1B (climate change effects on the management action and implications for climate-smart design) for 12 management actions for the Guánica Bay watershed and associated coral reefs.

S4 Table. Second phase output for Adaptation Design Tool Worksheets 1A (effect of climate change on stressors) and 1B (effects of climate change on management actions and how to adapt management actions) for three management actions for the Guánica Bay
watershed and associated coral reefs.

S5 Table. Five-column summaries of Adaptation Design Tool Worksheets 1A and 1B (effect of climate change on stressors and effects of climate change on management actions and how to adapt management actions, respectively) for 12 management actions for the Guánica Bay watershed and associated coral reefs.

S1 File. R script for conducting the southwest Puerto Rico resilience assessment.

S1 Data. Inputs and outputs for southwest Puerto Rico coral reef resilience assessment. Spreadsheet includes raw indicator values, rescaled indicator values, stressor values, resilience ranks and quartiles, and more. These data were used to make Fig 4.

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Author Contributions

Conceptualization: David A. Gibbs, Jordan M. West, Patricia Bradley.
Data curation: David A. Gibbs, Jordan M. West.
Formal analysis: David A. Gibbs.
Funding acquisition: Jordan M. West.
Investigation: David A. Gibbs, Jordan M. West, Patricia Bradley.
Methodology: David A. Gibbs, Jordan M. West, Patricia Bradley.
Project administration: David A. Gibbs, Jordan M. West, Patricia Bradley.
Resources: David A. Gibbs, Jordan M. West.
Software: David A. Gibbs.
Supervision: Jordan M. West.
Visualization: David A. Gibbs, Jordan M. West.
Writing – original draft: David A. Gibbs, Jordan M. West, Patricia Bradley.
Writing – review & editing: David A. Gibbs, Jordan M. West, Patricia Bradley.

References

1. CCSP, Baron JS, Griffith B, Joyce LA, Kareiva P, Keller BD, et al. Preliminary review of adaptation options for climate sensitive ecosystems and resources (Sap 4.4). In: Julius SH, West JM, editors. A report by the US climate change science program and the subcommittee on global change research. Washington, D.C.: US Environmental Protection Agency; 2008. p. 873.

2. Hernández-Delgado E, Shivlani M, Sabat A. Ecosystem-Based and Community-Based Model Integration to Designate Coral Reef No-Take Marine Protected Areas: A Case Study from Puerto Rico. Natural Resources. 2014; 5: 538–560. https://doi.org/10.4236/nr.2014.510049

3. Lipton D, Rubenstein MA, Weiskopf SR, Carter S, Peterson J, Crozier L, et al. Ecosystem, Ecosystem Services, and Biodiversity. Impacts, Risks, and Adaptation in the United States: Fourth National Climate Assessment. Volume II Reimkeller DR, Avery CW, Easterling DR, Kunkel KE, Lewis KLM, Maycock TK, and Stewart BC (eds). Washington, D.C.: U.S. Global Change Research Program; 2018. pp. 268–321. https://doi.org/10.7930/NCA4.2018.CH7

4. Donovan MK, Adam TC, Shantz AA, Speare KE, Munsterman KS, Rice MM, et al. Nitrogen pollution interacts with heat stress to increase coral bleaching across the seascape. Proc Natl Acad Sci USA. 2020; 117: 5351. https://doi.org/10.1073/pnas.1915395117 PMID: 32094188

5. Stein BA, Glick P, Edelson N, Staudt A. Climate-Smart Conservation: Putting Adaptation Principles into Practice. Washington, D.C.: National Wildlife Federation; 2014.

6. Stein BA, Glick P, Edelson N, Staudt A. Climate-Smart Conservation: Putting Adaptation Principles into Practice. Washington, D.C.: National Wildlife Federation; 2014. pp. 87–108.

7. Wagner D, Polhemus D. Climate Change Vulnerability Assessment for the Papahānaumokuākea Marine National Monument. Silver Spring, MD: U.S. Department of Commerce, National Oceanic and Atmospheric Administration, Office of National Marine Sanctuaries; 2016. Report No.: Marine Sanctuaries Conservation Series ONMS-16-03.

8. Maynard JA, Lewis K, Brown J, Ahmadia G. Assessing the relative resilience of the coral reefs of St. Croix, USVI. The Nature Conservancy and NOAA Coral Reef Conservation Program; 2014.

9. Grossarth SK, Hecht AD. Sustainability at the U.S. Environmental Protection Agency: 1970–2020. Ecological Engineering. 2007; 30: 1–8. https://doi.org/10.1016/j.ecoleeng.2006.09.011

10. McCleanah TR, Donner SD, Maynard JA, MacNeil MA, Graham NAJ, Maina J, et al. Prioritizing Resilient Indicators to Support Coral Reef Management in a Changing Climate. PLoS One. 2012; 7: e42884. https://doi.org/10.1371/journal.pone.0042884 PMID: 22926168

11. Gross J, Johnson K, Glick P, Hall K. Chapter 6. Understanding Climate Change Impacts and Vulnerability. Climate-Smart Conservation: Putting Adaptation Principles into Practice. Washington, D.C.: National Wildlife Federation; 2014. pp. 102–117. https://doi.org/10.1007/s00267-016-0774-3 PMID: 27734086
20. Smith A, Yee SH, Russell M, Awkerman J, Fisher WS. Linking ecosystem service supply to stakeholder concerns on both land and sea: An example from Guánica Bay watershed, Puerto Rico. Ecological Indicators. 2017; 74: 371–383. https://doi.org/10.1016/j.ecolind.2016.11.036

21. CWP C for WP. Guánica Bay Watershed Management Plan. 2008.

22. Bradley MP, Dyson B, Fisher W, Rehr A. Coral Reef and Coastal Ecosystems Decision Support Workshop, April 27–29, 2010, Caribbean Coral Reef Institute, La Parguera, Puerto Rico. National Health and Environmental Effects Research Laboratory, Office of Research and Development, EPA; 2014.

23. Bradley MP, Fisher W, Dyson B, Yee S, Carriger J, Gambrirazzo G, et al. Application of a structured decision process for informing watershed management options in Guánica Bay, Puerto Rico. Narrangansett, RI: National Health and Environmental Effects Research Laboratory, Office of Research and Development, EPA; 2015. Report No.: EPA/600/R-15/248.

24. NMFS NMFS. Recovery Plan for Elkhorn (Acropora palmata) and Staghorn (A. cervicornis) Corals. Silver Spring, MD: Prepared by the Acropora Recovery Team for the National Marine Fisheries Service; 2015.

25. McKay L, Bonelid T, Dewald T, Johnston J, Moore R, Rea A. NHDPlus Version 2: User Guide. 2012.

26. Bauer L, Edwards K, Roberson KKW, Kendall M, Tormey S, Battista T. Shallow-water benthic habitats of Southwest Puerto Rico. Silver Spring, MD: NOAA NOS NCCOS; 2012. Report No.: 155.

27. Wilkinson C. Status of Coral Reefs of the World: 2004. Australian Institute of Marine Science; 2004.

28. García JR, Morelock J, Castro R, Goenaga C, Hernández-Delgado E. Puerto Rican reefs: research synthesis, present threats and management perspectives. In: Cortés J, editor. Latin American Coral Reefs. Amsterdam: Elsevier Science; 2003. pp. 111–130. https://doi.org/10.1016/B978-044451388-5/50006-0

29. García-Sais JR, Castro R, Sabater-Clavell J, Esteves R, Carlo M. Monitoring of coral reef communities at Isla Desecheo, Rincon, Mayaguez Bay, Guanica, Ponce and Isla Caja de Muerto, Puerto Rico, 2005. San Juan, PR: Final Report submitted to the Department of Natural and Environmental Resources of Puerto Rico; 2005 p. 131.

30. García-Sais JR, Appeldoorn R, Bruckner AW, Caldow C, Christensen JD, Lylstrom C, et al. The state of coral reef ecosystems of the commonwealth of Puerto Rico. Silver Spring, MD: National Oceanographic and Atmospheric Administration; 2005 pp. 95–105. Report No.: NOAA Technical Memorandum NOS NCCOS 11.

31. Morelock J, Ramírez WR, Bruckner AW, Carlo M. Status of coral reefs, southwest Puerto Rico. Caribbean Journal of Science Special Publication. 2001; 4: 57.

32. Torres JL. Impacts of sedimentation on the growth rates of Montastraea annularis in southwest Puerto Rico. Bulletin of Marine Science. 2001; 69: 631–637.

33. Otero E. Spatial and temporal patterns of water quality indicators in reef systems of southwestern Puerto Rico. Caribbean Journal of Science. 2009; 45: 168–180. https://doi.org/10.18475/cjso.v45i2.a5

34. Oliver LM, Fisher WS, Dittmar J, Hallock P, Campbell J, Quarles RL, et al. Contrasting responses of coral reef fauna and foraminiferal assemblages to human influence in La Parguera, Puerto Rico. Marine Environmental Research. 2014; 99: 95–105. https://doi.org/10.1016/j.marenvres.2014.04.005 PMID: 24840256

35. Velázquez-Domínguez A, Weil E, Bruckner AW. Climate Change and Coral Bleaching in Puerto Rico: Efforts and Challenges. 2003; Oahu, Hawai'i.

36. Viqueira Ríos RA. Implementation of the Guánica Bay Watershed Management Plan: Integrated Watershed Management Actions, Final Report 2012–2017. Protectores de Cuencas; 2018.

37. USCRRTF. Coral Reef Protected Areas: A Guide for Management. Washington, D.C.: U.S. Coral Reef Task Force Working Group on Ecosystem Science and Conservation, Department of Interior; 2000 p. 14.

38. Fabricius KE. Effects of terrestrial runoff on the ecology of corals and coral reefs: review and synthesis. Marine Pollution Bulletin. 2005; 50: 125–146. https://doi.org/10.1016/j.marpolbul.2004.11.028 PMID: 15737355

39. Sturmi P, Viqueira Ríos R, Ferguson R, Moore T. Addressing land based sources of pollution in Guánica, Puerto Rico. Proceedings of the 12th International Coral Reef Symposium, Cairns, Australia, 9–13 July 2012. Cairns, Australia; 2012.

40. Ramos-Scharrón CE, Thomaz EL. Runoff Development and Soil Erosion in a Wet Tropical Montane Setting under Coffee Cultivation. Land Degradation & Development. 2017; 28: 936–945. https://doi.org/10.1002/ldr.2567
41. PRCCC. Puerto Rico’s State of the Climate 2010–2013: Assessing Puerto Rico’s Social-Ecological Vulnerabilities in a Changing Climate. San Juan, PR: Puerto Rico Coastal Zone Management Program, Department of Natural and Environmental Resources, NOAA Office of Ocean and Coastal Resource Management; 2013.

42. U.S. Coral Triangle Initiative Support Program. Climate Change Adaptation for Coral Triangle Communities: Guide for Vulnerability Assessment and Local Early Action Planning (LEAP Guide). 2013 p. Prepared with support from the United States Agency for International Development.

43. Ramos-Scharrón CE. Sediment production from natural and disturbed surfaces in dry tropical areas of the Eastern Caribbean. Proceedings of the 7th Caribbean Island Water Resources Congress. St. Croix, USVI: Univ. of the Virgin Islands Water Resources Research Institute; 2007. pp. 26–30.

44. Ramos-Scharrón CE. The Effects of Land Development on Sediment Loading Rates into the Coastal Waters of the Islands of Culebra and Vieques, Puerto Rico. San Juan, PR: Submitted to Puerto Rico Department of Natural and Environmental Resources, Coastal Zone Management Program; 2009 p. 94.

45. Ramos-Scharron C. Sediment Production from Unpaved Roads in A Sub-Tropical Dry Setting—Southwestern Puerto Rico. CATENA. 2010; 82: 146–158. https://doi.org/10.1016/j.catena.2010.06.001

46. Maynard JA, McLeod E. How-to-guide for conducting resilience assessments. The Nature Conservancy; 2012.

47. Shivlani M, Koeneke R. Spatial characterization of artisanal fisheries in Puerto Rico: Geographic Information Systems (GIS) approach for assessing the regional effort and landings. 2011. pp. 61–66.

48. Eslinger DL, Carter HJ, Pendleton M, Burkhalter S, Allen M. OpenNSPECT: The Open-source Nonpoint Source Pollution and Erosion Comparison Tool. 2012.

49. Whitall D, Bauer L, Edwards K, Palt T, Caldwell C, Sherman C. Baseline Assessment of Guánica Bay, Puerto Rico in Support of Watershed Restoration. Silver Spring, MD: Prepared by the NCCOS Center for Coastal Monitoring and Assessment Biogeography Branch; 2013 p. 169. Report No.: NOAA Technical Memorandum NOS NCCOS 176.

50. Miller G, Lugo A. Guide to the ecological systems of Puerto Rico. San Juan, PR: U.S. Department of Agriculture, Forest Service, International Institute of Tropical Forestry; 2009 p. 437. Report No.: Gen. Tech. Rep. IITF-GTR-35.

51. Turner MG, Dale VH, Gardner RH. Predicting across scales: Theory development and testing. Landscape Ecology. 1989; 3: 245–252. https://doi.org/10.1007/BF00131542

52. Levin SA. The Problem of Pattern and Scale in Ecology: The Robert H. MacArthur Award Lecture. Ecology. 1992; 73: 1943–1967. https://doi.org/10.2307/1941447

53. Sayre NF. Ecological and geographical scale: parallels and potential for integration. Progress in Human Geography. 2005; 29: 276–290. https://doi.org/10.1111/j.0309132505ph546oa

54. Wiens JA. Spatial Scaling in Ecology. Functional Ecology. 1989; 3: 385–397. https://doi.org/10.2307/2389612