Residential exposure to air pollution, access to neighborhood greenspace and their association with hair cortisol concentrations in the second and third trimester of pregnancy

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Abstract

Background: Exposure to air pollution during pregnancy has been associated with adverse pregnancy outcomes in studies worldwide, other studies have described beneficial effects of residential greenspace on pregnancy outcomes. The biological mechanisms that underlie these associations are incompletely understood. Recent studies have shown that a biological stress response, with release of cortisol, may underlie associations between air pollution and health effects. The available research on air pollution exposure in relation to biological stress during pregnancy is still scarce.

Methods: We explored associations between residential exposure to air pollution, access to neighborhood greenspace and hair cortisol concentrations in a prospective pregnancy cohort study. We modelled participants’ residential air pollutant concentrations (particulate matter with an aerodynamic diameter $≤ 2.5 \mu m$, nitrogen dioxide (NO$_2$), black carbon (BC)), assessed residential distance to a major road and access to a neighborhood greenspace. Hair cortisol concentrations, reflecting cortisol secretion over a period of 3 months prior to sampling, were determined at the end of the second ($n = 133$) and third pregnancy trimester ($n = 81$).

Results: Three-month mean residential NO$_2$ and BC concentrations were positively associated with third pregnancy trimester hair cortisol concentrations. The residential distance to a major road was negatively associated with second and third trimester hair cortisol concentrations. Access to a greenspace of 10 hectares or more within 800 meters travel distance significantly moderated the association between residential proximity to a major road and second trimester hair cortisol concentrations. At an average residential distance of 304 meters from a major road, mean second trimester HCC were estimated 22% lower for mothers with access to a neighborhood greenspace (3.71 (95% CI: 3.24, 4.24) pg/mg hair) compared to mothers without access (4.22 (95% CI: 3.26, 5.47) pg/mg hair). The moderation tended towards significance in the third trimester ($p < 0.10$).

Conclusions: Increased residential exposure to air pollution and closer proximity to a major road are associated with an increased biological stress response in the second and third trimester of pregnancy, access to neighborhood greenspace may moderate the association.

Trial registration: The IPANEMA study is registered under number NCT02592005 at clinicaltrials.gov.

Background

In the past decade, epidemiological studies throughout the world have linked maternal exposure to road traffic and air pollution to adverse pregnancy outcomes such as low birth weight, preterm birth and intrauterine growth restriction [1–3]. These adverse birth outcomes not only increase perinatal morbidity and mortality, but also increase susceptibility to obesity, diabetes and cardiovascular diseases later in life [4,5]. In terms of maternal health, exposure to air pollution has been linked to hypertensive pregnancy disorders (gestational hypertension, pre-eclampsia) and gestational diabetes mellitus [6–8]. Both conditions amplify young women's risk of developing cardiovascular diseases later in life [9]. The adverse
impact of maternal exposure to air pollution on birth outcomes is of major public health importance, considering the ubiquitous nature of air pollution in urban settings [10]. The biological pathways that relate maternal exposure to air pollution to adverse pregnancy outcomes however, remain incompletely understood. Recent experimental animal research has shown that a neuroendocrine stress response is among the early biological responses triggered by exposure to fine particulate matter with an aerodynamic diameter $\leq 2.5 \mu m$ (PM$_{2.5}$) and exposure to nitrogen dioxide (NO$_2$) [11]. The biological stress response includes activation of the hypothalamic-pituitary-adrenal (HPA) axis and release of glucocorticoid stress hormones, with the glucocorticoid cortisol as its main downstream effector in humans [12]. The relevance of these experimental observations to humans has been confirmed in a few recent studies [13,14]. To date, most human studies have assessed short-term variations in cortisol levels in relation to air pollution exposure, using blood and saliva as a matrix. Longer-term cortisol concentrations are difficult to evaluate using blood and saliva, due to circadian variations in cortisol secretion and the need for multiple sampling [15]. Repeated sampling increases discomfort for study participants. Hair however, is a suitable matrix for the assessment of longer-term cortisol concentrations [16]. As cortisol is incorporated into growing hair, hair cortisol concentrations (HCC) retrospectively reflect cortisol secretion over a period of several months [17]. Importantly, chronic activation of the maternal HPA axis during pregnancy is associated with gestational hypertensive disorders, intrauterine growth restriction and developmental programming of disease susceptibility [18,19]. With regard to birth outcomes, a significant negative association between maternal HCC in the 2nd pregnancy trimester and gestational age at delivery has been reported [20].

Interestingly, a growing body of research suggests a beneficial relationship between residential access to greenspaces (publicly accessible vegetation, urban parks, forests) and pregnancy outcomes for both mothers and babies [21]. Residential exposure to air pollution and greenspace occur simultaneously, therefore, the adverse effects of air pollution may, to some degree, be moderated by the beneficial effects of residential access to greenspace [22].

To our knowledge, no research is available on residential air pollution exposure or access to neighborhood greenspace in relation to longer-term biological stress during pregnancy. Accordingly, the main aim of this study was to explore associations between residential exposure to air pollution and road traffic and maternal hair cortisol concentrations in a pregnancy cohort. In addition, we aimed to explore whether residential access to neighborhood greenspace could moderate associations between residential air pollution exposure and maternal biological stress during pregnancy.

**Methods**

**Study population and design**

This study was conducted in the framework of IPANEMA (Impact of Particulate Matter on Mothers and Babies in Antwerp), a prospective pregnancy cohort study of the Antwerp University Hospital (UZA) in collaboration with the Flemish Institute for Technological Research (VITO) and the University of Antwerp.
Pregnant women were recruited between April 2015 and January 2018 at the UZA prenatal clinic by a midwife or obstetrician at a gestational age of 12 to 14 weeks. The inclusion criteria were: a singleton pregnancy; the ability to fill out extensive Dutch questionnaires; delivery planned in the Antwerp University Hospital. All participating mothers gave written informed consent. The study protocol was approved by the ethical committee of the University of Antwerp (14/40/411) and registered under number NCT02592005 at clinicaltrials.gov. Health-related information on mothers and babies was extracted from the hospital records and questionnaires that participants completed at enrolment, during pregnancy and after delivery. These questionnaires provided detailed information on participants’ socio-demographic and lifestyle characteristics. A detailed protocol of the IPANEMA study can be found elsewhere [23].

Hair sample collection and cortisol measurement

At the end of second trimester of pregnancy in-hospital consultation and shortly after delivery, a strand of hair of at least 2mm thick was bound together with a cotton thread and cut close to the scalp from the posterior vertex region of the head. This area of the scalp exhibits the lowest intra-individual variability in HCC [24]. Hair samples were stored in paper envelopes at room temperature until analysis. When protected from ultraviolet light, cortisol concentrations in hair samples remain stable at room temperature for several years [25]. Cortisol concentrations were determined from the 3 cm of hair closest to the scalp. Based on an average hair growth of 1 cm per month, this length represents cortisol secretion in a 3-month period, a trimester, prior to sampling [26]. There is a wide consensus that the first 5–6 cm of hair nearest to a person’s scalp can reliably reflect HPA activity [27]. Analysis was performed at the Institute of Public Health, Department of Environmental Medicine of the University of Southern Denmark (SDU), using liquid chromatography and tandem mass spectrometry (LC-MS/MS) as described by Chen et al. [28], after minor modifications. Hair samples were washed with methanol and dried at room temperature. The 3 cm of hair closest to the scalp was cut into segments of 2-3 mm. A typical amount of hair weighed 20-30 mg. Aliquots of 100 mL 20 ng/mL isotope labeled cortisol (cortisol-D₄) were added as internal standard, together with 0.9 mL methanol. Samples were incubated in the dark at 25°C while whirl mixed at 2000 revolutions per minute for 5 days and subsequently centrifuged at 3000g for 5 minutes. 20 μL of the supernatant was injected onto a High-Performance Liquid Chromatography (HPLC) column. HPLC was performed using an Accella 1250 pump (Thermo Scientific, San Jose, CA) and a PAL autosampler (CTC analytics, Zwingen, Switzerland). The analytical column was a Kinetex C18 column, 100 x 4.6 mm (2.6 μm) equipped with a 2 x 4 mm C18 SecurityGuard column (Phenomenex, Torrance, CA). Isocratic elution was performed with a mobile phase system consisting of methanol and 0.1 M formic acid (80:20) at a flow rate of 400 μL/min for 6 min. After the peaks were eluted, a wash procedure was performed before the next samples was injected onto the column. The triple quadrupole mass spectrometer utilized was a TSQ Vantage (Thermo Scientific, San Jose, CA). The calibration curve and calculation of the sample concentration were based on the area ratio of the analyte/isotope labeled internal standard. Quality control samples were included in each series of samples. The limit of quantification (LOQ) for cortisol was 1.0 pg/mg hair. The intra-day repeatability coefficient of variation was 8.7% and the inter-day reproducibility coefficient of variation was 9.5%
Residential exposure assessment

Assessment of all residential exposure variables was based on the participants’ geocoded home address. Geographical Information System (GIS) analyses were carried out using ESRI ArcGIS software version 10.4 (Environmental Systems Research Institute, Redlands, California, USA). The residential degree of urbanization was assessed according to the Eurostat definition that classifies local administrative units as cities, towns, suburbs or rural areas based on a combination of geographical contiguity and population density, applied to 1 km² population grid cells [29]. We assessed residential exposure to fine particulate matter (PM$_{2.5}$), nitrogen dioxide (NO$_2$) and black carbon (BC), primary constituents of traffic-related air pollution. Residential exposure to PM$_{2.5}$, NO$_2$ and BC was modelled using a spatial temporal interpolation method. In Flanders, atmospheric pollutants are continuously measured by a network of automatic monitoring stations by the Flemish Environment Agency. The Belgian Interregional Environment Agency (IRCEL, Intergewestelijke Cel voor het Leefmilieu) uses these measurements together with information on land cover to interpolate the air pollutant concentrations on a 4x4 km² resolution [30]. These background results are combined with a bi-gaussian dispersion model based on emissions from point sources and line sources, the Immission Frequency Distribution Model (IFDM). The combined RIO-IFDM model chain produces daily averaged pollutant concentrations in Belgium on a high resolution receptor grid [31]. We calculated mean air pollutant concentrations at the residential address over a 3-month period before sampling, similar to the period of cortisol accumulation in the hair samples, and over a 1-year period before sampling. Residential proximity to major roads is often used as a surrogate measure of long-term exposure to traffic-related air pollution [32]. We calculated the straight-line distance from each residence to the nearest major road. Major roads included international motorways (E-roads) and the network of large national and local roads of Belgium (N-roads).

Residential access to a neighborhood greenspace was based on the 2016 version of the land-use map of Flanders, which maps land cover types, i.e. natural vegetated land cover and urban greenery, in 10x10 m² raster cells [33]. Green cells were clustered to assess the area and public accessibility of greenspace in the maternal residential surroundings. Access to a small neighborhood greenspace (NHGS) was defined as access to at least 0.2 hectares (ha) of greenspace within a travel distance of 400 meters (m) from residence, access to a large greenspace was defined as access to at least 10 ha of greenspace within a travel distance of 800 m from residence. In the large greenspace typology, small water bodies are included when surrounded by > 50% greenspace, agricultural land is included when surrounded by > 30% greenspace. More technical background information on the green typology indicators can be found elsewhere [34].

Potential covariates of hair cortisol concentrations

Possible covariates of HCC were identified based on available data within the IPANEMA cohort and on existing literature [35]. Tested covariates included maternal age, parity, maternal socioeconomic status (SES) defined as the highest educational attainment of the mother and categorized as low/intermediate/high, pre-existing chronic diseases (diabetes, asthma, cardiovascular disease), pre-
pregnancy body mass index (BMI), gestational week at sampling, season of sampling, smoking and alcohol consumption before pregnancy, systemic use of glucocorticoids and daily hair washing. We assessed maternal ethnic background as European/non-European country of birth since hair growth rate may be influenced by ethnicity [36]. We additionally tested variables that may have a link with both residential environment and biological stress, i.e. neighborhood SES and residential exposure to noise. A systematic review in the World Health Organization (WHO) European Region showed that lower neighborhood SES is usually linked with higher levels of air pollutants [37]. Independent from higher levels of exposure, deprived mothers may have a higher vulnerability, leading to more pronounced adverse health effects of a given environmental exposure [38]. The Area Deprivation Index (ADI) is a yearly calculated indicator for neighborhood SES on a sub-municipality level in Flanders (Statistics Flanders, n.d.). Deprivation is recorded by the Flemish Child and Family Government Agency (www.kindengezin.be). Selection criteria for deprivation are the family’s monthly income, the parents’ educational attainment, the children’s development, the parents’ employment situation, housing and health. If a family fulfils three or more criteria, it is considered to be underprivileged (OECD, 2000). The index of year X (%), i.e. 2017, considers all children born in year X, X-1 and X-2 that live in deprived households in a given area in Flanders, divided by the total number of children born in the area during the same period. The ADI of the participants’ neighborhood was subsequently categorized into tertiles representing low, intermediate and high area deprivation across the range of ADI among all participants.

Residential proximity to major roads may also lead to elevated noise levels [41]. Residential noise exposure levels were assessed using the Flemish strategic noise map of 2016, which includes major road infrastructure as defined in the EU-guideline 2002/49/EG [42]. The strategic noise map expresses noise levels in L_{den}, the average sound level over a 24 hour period with a penalty of 5 dB added for evening hours and a penalty of 10 dB added for nighttime hours [43]. The WHO guideline for average noise exposure produced by road traffic is set at 53 decibels (dB) L_{den}, road traffic noise above this level has been associated with adverse health effects, including adverse birth outcomes [44]. Noise exposure was therefore evaluated binary as exposure to a noise level ≥ 53 dB L_{den}.

**Statistical analysis**

Statistical analysis was performed using SPSS Statistics (version 25; IBM, Armonk, NY, USA) and R (version 2018; R Foundation for Statistical Computing, Vienna, Austria). Descriptive statistics provide an overview of study population characteristics, residential exposure characteristics and geometric mean HCC concentration with 95% confidence interval. Air pollution variables were logarithmically transformed (ln-scale) because of skewed distributions, distance to major roads was logarithmically transformed to reflect the non-linear distance decay of traffic-related exposure to air pollutants [45].

Spearman rank correlations between residential exposures variables were assessed, since correlations of 0.9 or higher between exposure variables indicate strongly connected exposures that cannot be disentangled [46]. The outcome variable HCC was logarithmically transformed to obtain a normal distribution. For HCC below the LOQ of 1 pg/mg hair, a random imputation from a log-normal probability
distribution was performed where the mean was allowed to depend on observed values for hair cortisone concentrations that were measured simultaneously with cortisol, since both glucocorticoids were highly correlated \((p < 0.01, \text{Pearson's } r = 0.711 \text{ for } 2^{\text{nd}} \text{ trimester cortisol and cortisone}, p < 0.01, \text{Pearson's } r = 0.758 \text{ for } 3^{\text{rd}} \text{ trimester cortisol and cortisone})\). Linear regression models were used to analyze associations between 3-month mean air pollutant concentrations \((\text{PM}_{2.5}, \text{NO}_2, \text{BC})\), distance to major roads and access to greenspace as a predictor and \(2^{\text{nd}}\) and \(3^{\text{rd}}\) trimester Hair Cortisol Concentrations as an outcome. Given the limited number of study participants, we decided not to adjust for a set of \textit{a priori} selected covariates. The final regression models were only adjusted for significant covariates \((p < 0.05)\). All assumptions of linear regression were checked. To quantify the association, the estimated change in HCC \((\beta)\) with 95% confidence interval \((95\% \text{ CI})\) is presented for an increase in exposure from the 25\(^{\text{th}}\) to the 75\(^{\text{th}}\) percentile.

Effect modification by access the neighborhood greenspace was assessed by adding the interaction term of exposure to air pollution or distance to major roads and access to greenspace into the regression model. The level of significance for estimates was set at \(p < 0.05\).

We conducted several sensitivity analyses to evaluate our results. We tested 1-year mean air pollutant concentrations in relation to \(2^{\text{nd}}\) and \(3^{\text{rd}}\) trimester HCC to confirm the robustness of 3-month mean results. We additionally adjusted our \(2^{\text{nd}}\) trimester models for frequency of hair washing, a significant determinant of \(2^{\text{nd}}\) trimester HCC, independent of biological stress.

**Results**

Hair samples for cortisol analysis were provided by 152 participants. Characteristics of the study population are described in Table 1. We excluded 3 participants due to inexplicable high HCC values (> 3 times the interquartile range above the third quartile). As a result, 149 pregnant women were included in this study, of which 133 women donated a sample at the end of the \(2^{\text{nd}}\) trimester (week 26 ± 1.6) and 81 women shortly after delivery (week 39 ± 1.6), 65 women donated a sample twice. Almost half of the 149 mothers (48%) was aged between 26 and 30 years, 61% of participants were primigravid. Most of the study participants were of European origin (75%, 21% data missing), enjoyed higher education (57%, 23% data missing) and were employed prior to their pregnancy (72%, 23.5% data missing).

Residential characteristics are described in Table 2. Study participants lived in cities (38%), towns and suburbs (62%) in Flanders, none of the participants lived in a rural area. The mean ADI of our study population was 16.4\% (95\% CI: 14.6, 18.1) whereas the mean 2017 ADI for the study region Antwerp was 17.6\% (Statistics Flanders, n.d.). We tested the significance of the association between ADI as an area-level SES indicator and maternal educational attainment as a personal SES-indicator. We did not observe a significant association between neighborhood SES and personal SES (Spearman rank \(r = -0.074, p = 0.404\)). A small neighborhood greenspace was accessible for 94\% of participants, 76\% had residential access to a large greenspace. Three-month geometric mean \(\text{PM}_{2.5}\) was 11.61 (95\% CI: 11.06, 12.21)
µg/m^3 and 11.55 (95% CI: 10.95, 12.18) µg/m^3 for 2nd trimester and 3rd trimester sampling respectively. Geometric mean NO\textsubscript{2} concentrations 3 months before sampling was 23.03 (95% CI: 21.67, 24.47) µg/m^3 for the 2nd trimester and 23.19 (95% CI: 21.5, 24.98) µg/m^3 for the 3rd trimester, 3-month geometric mean BC concentration 1.13 (95% CI: 1.05, 1.21) µg/m^3 for the 2nd trimester and 1.17 (95% CI: 1.07, 1.28) µg/m^3 for the 3rd trimester. Noise exposure was covered by the Flemish strategic noise map for 144 participants, 31.5% of participating mothers was exposed to noise levels \( \geq \) 53 dB. Geometric mean 2nd trimester HCC was 3.94 (95% CI: 3.49, 4.45) pg/mg hair, geometric mean 3rd trimester HCC was 6.12 (95% CI: 4.96, 7.56) pg/mg hair. The coefficient of variance (CV) of 2nd trim HCC was 51.5%, CV of 3rd trimester HCC was 51.8%. Second and third trimester cortisol concentrations of participants that donated a hair sample twice were moderately correlated \((n = 65, p < 0.01, \text{Pearson's } r = 0.571)\).

Spearman rank correlations of residential exposure characteristics are presented in Table 3 (2nd trimester study population) and Table 4 (3rd trimester study population).

We observed strong positive correlations between 3-month mean air pollutant concentrations \((r \text{ ranged from 0.61 to 0.89})\). Distance to major roads was negatively correlated with NO\textsubscript{2} and BC concentrations \((r \text{ ranged from -0.24 and -0.37})\), but not with PM\textsubscript{2.5} concentrations.

Access to a neighborhood greenspace did not significantly correlate with air pollutants and distance to major roads in the 2nd trimester. In the 3rd trimester study population, we did find weak negative correlations between access to a large neighborhood greenspace and air pollutants \((r \text{ ranged from -0.28 to -0.31})\) and a weak positive association of access to a large neighborhood greenspace with distance to major road \((r = 0.26)\). The ADI was weakly positively correlated with NO\textsubscript{2}, BC and noise exposure above the WHO guideline \((r \text{ ranged from 0.21 to 0.46})\) and negatively correlated with distance to major roads and access to a large neighborhood greenspace \((r \text{ ranged from -0.18 to -0.25})\).

Season of sampling and daily hair washing were identified as significant covariates of 2nd trimester HCC, no significant covariates were identified for 3rd trimester HCC (see Table S1 for details). None of the participants reported the systemic use of glucocorticoids. Residential noise exposure above the WHO guideline \((\geq 53 \text{ dB } L\text{den})\) was not significantly associated with 2nd or 3rd trimester HCC \((p = 0.871, p = 0.190 \text{ respectively})\). Nor did we find significant associations between the ADI and 2nd or 3rd trimester HCC \((p = 0.661, p = 0.388 \text{ resp.})\).

Results of the associations between air pollution exposure, access to neighborhood greenspace and maternal biological stress are presented in Table 5. We found a significant negative association between 3-month mean PM\textsubscript{2.5} concentrations and 2nd trimester HCC in the unadjusted model \((p = 0.009)\), the association did not remain significant after adjustment for season of sampling \((p = 0.357)\). In the 3rd trimester, 3-month mean PM\textsubscript{2.5} concentrations were not significantly associated with HCC.
We did not observe significant associations between 3-month mean NO\textsubscript{2} and BC concentrations and 2\textsuperscript{nd} trimester HCC. We did observe a significant positive association between 3-month mean NO\textsubscript{2} concentrations and 3\textsuperscript{rd} trimester HCC ($p = 0.016$). For an increase of 3-month mean residential NO\textsubscript{2} concentrations from 18.35 µg/m\textsuperscript{3} (p25) to 30 µg/m\textsuperscript{3} (p75), an increase of 3\textsuperscript{rd} trimester HCC with a factor 1.42 (95% CI: 1.07, 1.88) was estimated. We also observed a significant positive association between 3-month mean BC concentrations and 3\textsuperscript{rd} trimester HCC ($p = 0.032$). For an increase of 3-month mean residential BC concentrations from 0.84 µg/m\textsuperscript{3} (p25) to 1.48 µg/m\textsuperscript{3} (p75), an increase of 3\textsuperscript{rd} trimester HCC with a factor 1.37 (95% CI: 1.03, 1.82) was estimated. Residential 3-month mean NO\textsubscript{2} concentrations explained 6.2% of the variations in 3\textsuperscript{rd} trimester HCC, 3-month mean BC concentrations explained 4.6%.

Residential distance to a major road was negatively associated with second trimester HCC ($p = 0.016$ unadjusted model, $p = 0.11$ adjusted model). For an increase in distance to a major road from 143 m (p25) to 642 m (p75), a change of 2\textsuperscript{nd} trimester HCC with a factor 0.82 (95% CI: 0.70, 0.95) was estimated in the adjusted model. The model explained 7.8% of the variation in 2\textsuperscript{nd} trimester HCC. Distance to a major road was also negatively associated with 3\textsuperscript{rd} trimester HCC ($p = 0.04$). For an increase of distance to a major road from 114 m (p25) to 598 m (p75), a change of 3\textsuperscript{rd} trimester HCC with a factor 0.74 (95% CI: 0.55, 0.99) was estimated. Distance to a major road explained 4% of the variations in 3\textsuperscript{rd} trimester HCC. Access to a small neighborhood greenspace tended towards a significant negative association with 2\textsuperscript{nd} trimester HCC ($p = 0.117$ unadjusted model, $p = 0.061$ adjusted model) and was not significantly associated with 3\textsuperscript{rd} trimester HCC ($p = 0.354$). Access to a large greenspace tended towards a significant negative association with 2\textsuperscript{nd} trimester HCC ($p = 0.073$ unadjusted model, $p = 0.095$ adjusted) and with 3\textsuperscript{rd} trimester HCC ($p = 0.062$).

We tested whether access to a neighborhood greenspace moderated the associations between air pollution exposure, proximity to major roads and maternal biological stress. We found no significant interaction between access to a small or large neighborhood greenspace and air pollution constituents in relation to 2\textsuperscript{nd} or 3\textsuperscript{rd} trimester HCC (see Table S2 for details). We did observe a significant interaction between access to a large neighborhood greenspace (NHGS) and distance to major roads in relation to 2\textsuperscript{nd} trimester HCC in both the unadjusted model, as presented in Figure 1, and the model adjusted for season of sampling ($p = 0.021$, $p = 0.034$ resp.). At an average residential distance of 304 meters from a major road, estimated mean 2\textsuperscript{nd} trimester HCC were 22% lower for mothers with access to a large neighborhood greenspace (3.71 (95% CI: 3.24, 4.24) pg/mg hair) compared to mothers without access (4.22 (95% CI: 3.26, 5.47) pg/mg hair). The adjusted interaction model explained 10.8% of the variations in 2\textsuperscript{nd} trimester HCC.

Figure 1 Interaction between distance to a major road and access to a large neighborhood greenspace in relation to 2nd trimester HCC
The interaction between access to a large neighborhood greenspace (NHGS) and distance to major roads in relation to 3rd trimester HCC tended towards significance \((p = 0.073)\), as presented in Figure 2. At an average residential distance of 264 m from a major road, the model estimated 19% lower mean 3rd trimester HCC for mothers with access to a large neighborhood greenspace (5.46 (95% CI: 4.30, 6.95) pg/mg hair) compared to mothers without access (6.75 (95% CI: 4.46, 10.22) pg/mg hair). The interaction model explained 7.9% of the variations in 3rd trimester HCC.

Figure 2 Interaction between distance to a major road and access to a large neighborhood greenspace in relation to 3rd trimester HCC

In a sensitivity analysis, we evaluated the significance of associations between 1-year mean PM\(_{2.5}\), NO\(_2\) and BC concentrations and HCC to reflect the participants’ longer-term residential exposure to air pollution. Results are presented in Table S3. Extending the exposure period did not change our results, we found significant positive associations between 1-year mean NO\(_2\) concentrations and 3rd trimester HCC \((p = 0.013)\) and between 1-year mean BC concentrations and 3rd trimester HCC \((p = 0.046)\). The robustness of our results was also evaluated by additional adjustment of our 2nd trimester model with daily hair washing, a significant but external, non-biological determinant of HCC. Our results, presented in Table S4 and S5, remained robust. Residential distance to a major road, adjusted for season of sampling and daily hair washing was significantly associated with 2nd trimester HCC \((n = 103, p = 0.006, R^2 = 17.4\%)\). In accordance, access to a large greenspace significantly moderated the association between distance to major roads and 2nd trimester HCC in the model adjusted for season of sampling and daily hair washing \((n = 103, p = 0.001, R^2 = 22.1\%)\). The model estimated 17.66% lower mean 2nd trimester HCC for participants with access to large NH greenspace compared to participants without access to large NH greenspace (2.89 (95% CI: 2.34, 3.57) versus 3.51 (95% CI: 2.54, 4.85) pg/mg hair).

**Discussion**

This study provides new insights in the relation between residential exposure to air pollution and road traffic and hair cortisol as a biomarker for longer-term biological stress during pregnancy. We observed significant positive associations between residential 3-month mean NO\(_2\) and BC concentrations and maternal biological stress in the 3rd pregnancy trimester. It should be noted that NO\(_2\) exposure levels were strongly correlated with BC exposure levels \((r = 0.89)\), making it impossible to disentangle the effects of both pollutants. In urban settings, road traffic is the principal source of NO\(_2\) and BC in ambient air [47]. We also observed significant associations between residential proximity to major roads and maternal biological stress in the 3rd pregnancy trimester. As previously reported, residential proximity to major roads and maternal biological stress in the 2nd pregnancy trimester were significantly associated [48]. The difference in significant associations between traffic-related exposures and 2nd and 3rd trimester HCC may be due to the difference in study population between both trimesters or to the increase in circulating cortisol concentrations towards the end of pregnancy, which is a normal biological process [20]. Our
observations are in line with recent human studies that reported associations between air pollutants and short-term variations in cortisol secretion. In a panel study among 43 students in Shanghai, residential exposure to PM$_{2.5}$ was associated with higher serum cortisol levels [13]. In a cross-sectional analysis of 1793 adults, residential NO$_2$ exposure was associated with higher wake-up salivary cortisol [14]. To our knowledge, only one epidemiological study has examined the association between personal air pollution exposure and HCC; the study, including Belgian schoolchildren and adolescents, did not find a significant relationship [49]. Pregnancy however, is a vulnerable period for both mother and fetus [50]. Several mechanisms potentially underlie the association between air pollution and biological stress during pregnancy. Air pollutants may induce oxidative stress and low grade inflammation [51]. Depending on size and chemical composition, inhaled air pollution constituents may translocate from the lungs to the systemic circulation or migrate via olfactory transport to the brain and directly interact with brain tissues including the hypothalamus [11]. During pregnancy, oxidative stress is known to be higher than in the non-pregnant state; residential exposure to air pollutants and road traffic may further amplify the level of maternal oxidative stress [52]. Oxidative stress may in turn lead to low grade inflammation and HPA axis activation, resulting in a marked increase in the secretion of cortisol into the circulation [53]. In addition to indirect activation of the HPA axis by systemic low grade inflammation, low grade inflammation in the brain may directly activate the hypothalamus [54]. Flanders, the IPANEMA study region, is characterized by a dense road network and high emissions from traffic [55]. The fraction of the Flemish population, living and working in close proximity to traffic, is high and access to neighborhood greenspace is typically limited. Interestingly, in our urban and suburban pregnancy cohort, access to significantly moderated the association between residential proximity to traffic and maternal biological stress in the 2$^{nd}$ pregnancy trimester. Beneficial relationships between residential access to greenspace and hair cortisol concentrations have been described in previous studies [56,57]; whereas other studies have reported a beneficial impact of surrounding greenness on fetal growth and birth weight [58–61]. Neighborhood greenspace may improve health by relieving psychophysiological stress, supporting physical activity, increasing social contacts and by reducing exposure to air pollution, noise and excessive heat [62–65]. In our study, we found a weak inverse correlation between access to a large greenspace and residential air pollutant concentrations in the 3$^{rd}$ pregnancy trimester, but not in the 2$^{nd}$ trimester. This suggest a moderating effect of residential access to a large greenspace, independent of the effect on air pollution exposure levels.

The added value of prospective cohort studies such as IPANEMA, is the possibility to provide more insight into early pathophysiological mechanisms, triggered by air pollution exposure, in real-world settings. The urban and suburban character of the IPANEMA cohort made it possible to go deeper into traffic-related air pollution exposure, notwithstanding the low number of participants. Residential exposure to air pollution was estimated using a high spatial resolution model, residential mobility of participants was considered. In addition to maternal traffic and air pollution exposure, we took access to neighborhood greenspaces into account. Evaluation of only one residential environmental exposure i.e. air pollution, ignoring potential interaction with other jointly occurring exposures i.e. access to greenspaces, could lead to an inaccurate estimate of the true effect of exposures [66]. We measured hair cortisol concentrations, a novel
method in epidemiological studies to retrospectively determine longer-term biological stress in a non-invasive and reliable way. Hair samples were collected according to a strict protocol by trained midwives at the in-hospital consultation to avoid interindividual differences in hair collection. Some limitations of the study need to be addressed. In this study, we had a limited number of participants and did not have the same study population in the second and third pregnancy trimester, leading to differences in residential exposures. We had a considerable percentage of missing questionnaire-based data. Future prospective cohort studies, ideally including a larger number of participants from pre-conception onwards, should enhance efforts to collect questionnaires from all participants, including relevant information on time spent in residential greenspaces, physical activity, wellbeing and health. The exposure assessment was limited to residential surroundings, we did not consider air pollution exposure while commuting and working. IPANEMA participants were mostly of a higher socioeconomic status, it was therefore not possible to explore increased vulnerability to environmental exposure in participants with lower SES. Moreover, the neighborhood SES indicator in this study did not reflect participants with lower socioeconomic status. In literature, the pattern of air pollution is often described as U-shaped, although the most deprived areas have the highest levels of poor air quality, the least deprived areas also experience higher levels of air pollutants than some other social groups [37]. In our cohort, the mean ADI was 18.1% for lower educated mothers, 15.3% for medium education mothers and, as described in literature, we observed a slightly higher ADI of 15.6% for higher educated women. Future studies should enhance efforts to include participants of all SES.

Conclusions

This study observed significant positive associations between residential exposure to traffic-related air pollution during pregnancy and longer-term biological stress in the 2nd and 3rd trimester of pregnancy. In the 2nd trimester of pregnancy, the association was significantly moderated by residential access to a large neighborhood greenspace. Air pollution and urban spatial planning are in the center of public debate in Flanders. Because of the ubiquitous nature of traffic-related air pollution and the adverse pregnancy outcomes that have been associated with elevated maternal biological stress for both mothers and babies, even a small increase in maternal biological stress may be of public health interest. Our research, if confirmed in future studies, may provide guidance towards a more sustainable urban planning and support environmental health protection for both pregnant women and their babies.

Abbreviations
Declarations

Ethics approval and consent to participate

The study protocol was approved by the ethical committee of the University of Antwerp (14/40/411). All participants gave written informed consent.

All co-authors have read the manuscript and consented with publication.

Consent for publication

Not applicable

Availability of data and materials
The datasets used and analyzed during the current study are available from the corresponding author on reasonable request.

**Competing interests**

The authors declare that they have no competing interests.

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**Authors' contributions**

VV contributed to the study design, data acquisition, analysis and interpretation and was the first author of the manuscript draft. SR contributed to data interpretation, data visualization and revision of the manuscript. NL contributed to the study concept and design and revision of the manuscript. EG and AC contributed to the data analysis and revision of the manuscript. LP, EV, WL and CV contributed to the acquisition of environmental data. PM contributed to the environmental data analysis. FN conducted the analysis of hair cortisol and therefore contributed to the biomarker data acquisition. LVdE contributed to the in-hospital data acquisition. YJ supervised the IPANEMA study and contributed to the study concept and acquisition of data. GS contributed to the study concept, study design and revision of the manuscript and provided general guidance.

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**References**

1. Kingsley SL, Eliot MN, Whitsel EA, Huang YT, Kelsey KT, Marsit CJ, et al. Maternal residential proximity to major roadways, birth weight, and placental DNA methylation. Environ Int. Elsevier Ltd; 2016;92–93:43–9.

2. Dadvand P, Ostro B, Figueras F, Foraster M, Basagaña X, Valentín A, et al. Residential proximity to major roads and term low birth weight: The roles of air pollution, heat, noise, and road-adjacent trees.
3. Zhu X, Liu Y, Chen Y, Yao C, Che Z, Cao J. Maternal exposure to fine particulate matter (PM2.5) and pregnancy outcomes: a meta-analysis. Environ Sci Pollut Res. Springer Verlag; 2015;22:3383–96.

4. Barker DJP, Godfrey KM, Gluckman PD, Harding JE, Owens JA, Robinson JS. Fetal nutrition and cardiovascular disease in adult life. Lancet. 1993;341:938–41.

5. Barouki R, Melén E, Herceg Z, Beckers J, Chen J, Karagas M, et al. Epigenetics as a mechanism linking developmental exposures to long-term toxicity. Environ. Int. Elsevier Ltd; 2018. p. 77–86.

6. Pedersen M, Stayner L, Slama R, Sørensen M, Figueras F, Nieuwenhuijsen MJ, et al. Ambient air pollution and pregnancy-induced hypertensive disorders: A systematic review and meta-analysis. Hypertension. Lippincott Williams and Wilkins; 2017;64:494–500.

7. Pedersen M, Halldorsson TI, Olsen SF, Hjortebjerg D, Ketzel M, Grandström C, et al. Impact of road traffic pollution on pre-eclampsia and pregnancy-induced hypertensive disorders. Epidemiology. Lippincott Williams and Wilkins; 2017;28:99–106.

8. Elshahidi MH. Outdoor air pollution and gestational diabetes mellitus: A systematic review and meta-analysis. Iran. J. Public Health. Iranian Journal of Public Health; 2019. p. 9–19.

9. Abramson BL, Melvin RG. Cardiovascular risk in women: Focus on hypertension. Can. J. Cardiol. Pulsus Group Inc.; 2014. p. 553–9.

10. Dadvand P, Parker J, Bell ML, Bonzini M, Brauer M, Darrow LA, et al. Maternal exposure to particulate air pollution and term birth weight: A multi-country evaluation of effect and heterogeneity. Environ Health Perspect. 2013;121:367–73.

11. Thomson EM. Air Pollution, Stress, and Allostatic Load: Linking Systemic and Central Nervous System Impacts. J Alzheimers Dis. IOS Press; 2019;69:597–614.

12. Thomson EM, Filiatreault A, Guénette J. Stress hormones as potential mediators of air pollutant effects on the brain: Rapid induction of glucocorticoid-responsive genes. Environ Res. Academic Press Inc.; 2019;178.

13. Niu Y, Chen R, Xia Y, Cai J, Ying Z, Lin Z, et al. Fine particulate matter constituents and stress hormones in the hypothalamus–pituitary–adrenal axis. Environ Int. Elsevier Ltd; 2018;119:186–92.

14. Hajat A, Hazlehurst MF, Golden SH, Merkin SS, Seeman T, Szpiro AA, et al. The cross-sectional and longitudinal association between air pollution and salivary cortisol: Evidence from the Multi-Ethnic Study of Atherosclerosis. Environ Int. Elsevier Ltd; 2019;131.

15. Lee DY, Kim E, Choi MH. Technical and clinical aspects of cortisol as a biochemical marker of chronic stress. BMB Rep. The Biochemical Society of the Republic of Korea; 2015. p. 209–16.

16. Stalder T, Kirschbaum C. Analysis of cortisol in hair - State of the art and future directions. Brain. Behav. Immun. Brain Behav Immun; 2012. p. 1019–29.

17. Kirschbaum C, Tietze A, Skoluda N, Dettenborn L. Hair as a retrospective calendar of cortisol production-Increased cortisol incorporation into hair in the third trimester of pregnancy. Psychoneuroendocrinology. Psychoneuroendocrinology; 2009;34:32–7.
18. Cottrell EC, Seckl JR, Holmes MC, Wyrwoll CS. Foetal and placental 11B-HSD2: A hub for developmental programming. Acta Physiol. 2014.

19. Bärebring L, O’Connell M, Winkvist A, Johannsson G, Augustin H. Serum cortisol and vitamin D status are independently associated with blood pressure in pregnancy. J Steroid Biochem Mol Biol. Elsevier Ltd; 2019;189:259–64.

20. Hoffman MC, Mazzoni SE, Wagner BD, Laudenslager ML, Ross RG. Measures of Maternal Stress and Mood in Relation to Preterm Birth. Obstet Gynecol. Lippincott Williams and Wilkins; 2016;127:545–52.

21. Banay RF, Bezold CP, James P, Hart JE, Laden F. Residential greenness: Current perspectives on its impact on maternal health and pregnancy outcomes. Int. J. Womens. Health. Dove Medical Press Ltd; 2017. p. 133–44.

22. Kim S, Kim H, Lee JT. Interactions between ambient air particles and greenness on cause-specific mortality in seven Korean metropolitan cities, 2008-2016. Int J Environ Res Public Health. MDPI AG; 2019;16.

23. Van Den Eeden L, Lambrechts N, Verheyen V, Berth M, Schoeters G, Jacquemyn Y. Impact of particulate matter on mothers and babies in Antwerp (IPANEMA): A prospective cohort study on the impact of pollutants and particulate matter in pregnancy. BMJ Open. 2018;8.

24. Sauvé B, Koren G, Walsh G, Tokmakejian S, Van Uum SHM. Measurement of cortisol in human hair as a biomarker of systemic exposure. Clin Investig Med. 2007;30.

25. Greff MJE, Levine JM, Abuzgaia AM, Elzagallaai AA, Rieder MJ, van Uum SHM. Hair cortisol analysis: An update on methodological considerations and clinical applications. Clin. Biochem. Elsevier Inc.; 2019. p. 1–9.

26. Abell JG, Stalder T, Ferrie JE, Shipley MJ, Kirschbaum C, Kivimäki M, et al. Assessing cortisol from hair samples in a large observational cohort: The Whitehall II study. Psychoneuroendocrinology. Elsevier Ltd; 2016;73:148–56.

27. Russell E, Kirschbaum C, Laudenslager ML, Stalder T, de Rijke Y, van Rossum EFC, et al. Toward Standardization of Hair Cortisol Measurement. Ther Drug Monit. 2015;37:71–5.

28. Chen Z, Li J, Zhang J, Xing X, Gao W, Lu Z, et al. Simultaneous determination of hair cortisol, cortisone and DHEAS with liquid chromatography-electrospray ionization-tandem mass spectrometry in negative mode. J Chromatogr B Anal Technol Biomed Life Sci. J Chromatogr B Analyt Technol Biomed Life Sci; 2013;929:187–94.

29. Eurostat. Urban Europe statistics on cities, town and suburbs. Luxembourg; 2016.

30. Janssen S, Dumont G, Fierens F, Mensink C. Spatial interpolation of air pollution measurements using CORINE land cover data. Atmos Environ. 2008;

31. Lefebvre W, Degrawe B, Beckx C, Vanhulsel M, Kochan B, Bellemans T, et al. Presentation and evaluation of an integrated model chain to respond to traffic- and health-related policy questions. Environ Model Softw. Elsevier; 2013;40:160–70.
32. Wilhelm M, Ritz B. Residential proximity to traffic and adverse birth outcomes in Los Angeles County, California, 1994-1996. Environ Health Perspect. Environ Health Perspect; 2003;111:207–16.

33. Poelmans L, Janssen L, Hambsch L. Landgebruik en ruimtebeslag in Vlaanderen, toestand 2016, uitgevoerd in opdracht van het Vlaams Planbureau voor Omgeving. 2019.

34. Verachtert E, Poelmans L, Vermeiren K, Hendrix R. Technische fiche groentypologieën Stadsmonitor. Studie uitgevoerd in opdracht van Agentschap Binnenlands Bestuur. 2018.

35. Stalder T, Steudte-Schmiedgen S, Alexander N, Klucken T, Vater A, Wichmann S, et al. Stress-related and basic determinants of hair cortisol in humans: A meta-analysis. Psychoneuroendocrinology. Elsevier Ltd; 2017. p. 261–74.

36. Bleker LS, Roseboom TJ, Vrijkotte TG, Reynolds RM, de Rooij SR. Determinants of cortisol during pregnancy – The ABCD cohort. Psychoneuroendocrinology. Elsevier Ltd; 2017;83:172–81.

37. Fairburn J, Schüle SA, Dreger S, Hilz LK, Bolte G. Social inequalities in exposure to ambient air pollution: A systematic review in the WHO European region. Int. J. Environ. Res. Public Health. 2019.

38. Schüle SA, Gabriel KMA, Bolte G. Relationship between neighbourhood socioeconomic position and neighbourhood public green space availability: An environmental inequality analysis in a large German city applying generalized linear models. Int J Hyg Environ Health. Elsevier GmbH; 2017;220:711–8.

39. Statistics Flanders [Internet]. Available from: https://www.statistiekvlaanderen.be/kansarmoede-index-van-kind-en-gezin

40. OECD Country Note Early Childhood Education and Care Policy in the Flemish Community of Belgium. 2000.

41. Moudon AV. Real Noise from the Urban Environment. How Ambient Community Noise Affects Health and What Can Be Done About It. Am J Prev Med. Am J Prev Med; 2009;37:167–71.

42. Eionet. Environmental Noise Directive.

43. Miedema HME, Vos H. Noise annoyance from stationary sources: Relationships with exposure metric day–evening–night level (DENL) and their confidence intervals. J Acoust Soc Am. Acoustical Society of America (ASA); 2004;116:334–43.

44. WHO. Environmental noise guidelines for the European Region. Copenhagen; 2018.

45. Weaver AM, Wellenius GA, Wu WC, Hickson DA, Kamalesh M, Wang Y. Residential distance to major roadways and cardiac structure in African Americans: Cross-sectional results from the Jackson Heart Study. Environ Heal A Glob Access Sci Source. BioMed Central Ltd.; 2017;16.

46. Robinson O, Tamayo I, de Castro M, Valentin A, Giorgis-Allemand L, Krog NH, et al. The urban exposome during pregnancy and its socioeconomic determinants. Environ Health Perspect. Public Health Services, US Dept of Health and Human Services; 2018;126.

47. Liu S V., Chen F lin, Xue J. A meta-analysis of selected near-road air pollutants based on concentration decay rates. Heliyon. Elsevier Ltd; 2019;5.
48. Verheyen V, Van den Eeden L, Lambrechts N, Remy S, Govarts E, Nielsen F, et al. Residential proximity to major roads and neighbourhood green space in relation to biological stress in the second trimester of pregnancy in the IPANEMA cohort. Environ Epidemiol. 2019;3:411–2.

49. Van Aart CJC, Michels N, Sioen I, De Decker A, Bijnens EM, Janssen BG, et al. Residential landscape as a predictor of psychosocial stress in the life course from childhood to adolescence. Environ Int. Elsevier Ltd; 2018;120:456–63.

50. Makri A, Stilianakis Ni. Vulnerability to air pollution health effects. Int J Hyg Environ Health. 2008;

51. Brook RD, Rajagopalan S, Pope CA, Brook JR, Bhatnagar A, Diez-Roux A V., et al. Particulate matter air pollution and cardiovascular disease: An update to the scientific statement from the american heart association. Circulation. 2010. p. 2331–78.

52. Miller MR, Shaw CA, Langrish JP. From particles to patients: Oxidative stress and the cardiovascular effects of air pollution. Future Cardiol. 2012. p. 577–602.

53. John CD, Buckingham JC. Cytokines: Regulation of the hypothalamo-pituitary-adrenocortical axis. Curr. Opin. Pharmacol. 2003. p. 78–84.

54. Fuertes E, Standl M, Forns J, Berdel D, Garcia-Aymerich J, Markevych I, et al. Traffic-related air pollution and hyperactivity/inattention, dyslexia and dyscalculia in adolescents of the German GINIplus and LISAplus birth cohorts. Environ Int. Elsevier Ltd; 2016;97:85–92.

55. De Prins S, Dons E, Van Poppel M, Int Panis L, de Mieroop E Van, Nelen V, et al. Airway oxidative stress and inflammation markers in exhaled breath from children are linked with exposure to black carbon. Environ Int. 2014;73.

56. Honold J, Lakes T, Beyer R, van der Meer E. Restoration in Urban Spaces. Environ Behav. SAGE PublicationsSage CA: Los Angeles, CA; 2016;48:796–825.

57. Gidlow CJ, Randall J, Gillman J, Smith GR, Jones M V. Natural environments and chronic stress measured by hair cortisol. Landsc Urban Plan. Elsevier; 2016;148:61–7.

58. Ebisu K, Holford TR, Bell ML. Association between greenness, urbanicity, and birth weight. Sci Total Environ. Elsevier B.V.; 2016;542:750–6.

59. Dadvand P, Sunyer J, Basagaña X, Ballester F, Lertxundi A, Fernández-Somoano A, et al. Surrounding greenness and pregnancy outcomes in four Spanish birth cohorts. Environ Health Perspect. Environ Health Perspect; 2012;120:1481–7.

60. Markevych I, Fuertes E, Tiesler CMT, Birk M, Bauer CP, Koletzko S, et al. Surrounding greenness and birth weight: Results from the GINIplus and LISAplus birth cohorts in Munich. Heal Place. Elsevier Ltd; 2014;26:39–46.

61. Nieuwenhuijsen MJ, Agier L, Basagaña X, Urquiza J, Tamayo-Uria I, Giorgis-Allemand L, et al. Influence of the urban exposome on birth weight. Environ Health Perspect. Public Health Services, US Dept of Health and Human Services; 2019;127.

62. Dadvand P, de Nazelle A, Figueras F, Basagaña X, Su J, Amoly E, et al. Green space, health inequality and pregnancy. Environ Int. Elsevier Ltd; 2012;40:110–5.
63. Ward Thompson C, Roe J, Aspinall P, Mitchell R, Clow A, Miller D. More green space is linked to less stress in deprived communities: Evidence from salivary cortisol patterns. Landsc Urban Plan. Elsevier; 2012;105:221–9.

64. Roe JJ, Aspinall PA, Thompson CW. Coping with stress in deprived urban neighborhoods: What is the role of green space according to life stage? Front Psychol. Frontiers Media S.A.; 2017;8.

65. Markevych I, Schoierer J, Hartig T, Chudnovsky A, Hystad P, Dzhambov AM, et al. Exploring pathways linking greenspace to health: Theoretical and methodological guidance. Environ. Res. Academic Press Inc.; 2017. p. 301–17.

66. Klompmaker JO, Janssen NAH, Bloemsma LD, Gehring U, Wijga AH, Brink C Vanden, et al. Associations of combined exposures to surrounding green, air pollution, and road traffic noise with cardiometabolic diseases. Environ Health Perspect. Public Health Services, US Dept of Health and Human Services; 2019;127.

Tables

Table 1. Basic characteristics of the study participants
| Characteristic                                   | n (%)  |
|-------------------------------------------------|--------|
| **Age**                                         |        |
| ≤ 25                                            | 17 (11.4) |
| 26-30                                           | 71 (47.7) |
| 31-35                                           | 44 (29.5) |
| >35                                             | 17 (11.4) |
| Missing                                         | 0      |
| **Parity**                                      |        |
| 0                                               | 91 (61.1) |
| 1                                               | 43 (28.9) |
| ≥ 2                                             | 15 (10) |
| Missing                                         | 0      |
| **Pre-existing chronic diseases**                |        |
| No                                              | 135 (90.6) |
| Yes                                             | 12 (8.1) |
| Missing                                         | 2 (1.3) |
| **Pre-pregnancy Body Mass Index (kg/m²)**        |        |
| Underweight (<18.5)                             | 8 (5.4) |
| Normal (18.5-24.9)                              | 84 (56.4) |
| Overweight (25-29.9)                            | 20 (13.4) |
| Obese (≥30)                                     | 12 (8.1) |
| Missing                                         | 25 (16.7) |
| **Smoking before pregnancy**                    |        |
| Never                                           | 94 (63.1) |
| Former smoker                                   | 21 (14.1) |
| Missing                                         | 34 (22.8) |
| **Alcohol consumption before pregnancy**        |        |
| No                                              | 16 (10.7) |
| Yes                                             | 99 (66.4) |
| Missing                                         | 34 (22.8) |
| **Ethnic background**                           |        |
| European          | 112 (75.2) |
|-------------------|-----------|
| Non-European      | 4 (2.7)   |
| Missing           | 33 (21.1) |

**Educational attainment**

| Level                | Count (Percentage) |
|----------------------|--------------------|
| Low (Basic level)    | 15 (10.1)          |
| Intermediate (Secondary school) | 15 (10.1)        |
| High (Higher education) | 85 (57.0)        |
| Missing              | 34 (22.8)          |

**Pre-pregnancy employment**

| Status | Count (Percentage) |
|--------|--------------------|
| No     | 7 (4.7)            |
| Yes    | 107 (71.8)         |
| Missing| 35 (23.5)          |

**Daily hair washing**

| Status | Count (Percentage) |
|--------|--------------------|
| No     | 97 (65.1)          |
| Yes    | 18 (12.1)          |
| Missing| 34 (22.8)          |

**Season of 2nd trimester sampling (n=133)**

| Season | Count (Percentage) |
|--------|--------------------|
| Autumn | 28 (21.2)          |
| Winter | 23 (17.3)          |
| Spring | 38 (28.6)          |
| Summer | 44 (33.1)          |
| Missing| 0                  |

**Season of 3rd trimester sampling (n=81)**

| Season | Count (Percentage) |
|--------|--------------------|
| Autumn | 23 (28.4)          |
| Winter | 21 (25.9)          |
| Spring | 12 (14.8)          |
| Summer | 25 (30.9)          |
| Missing| 0                  |

Note: pre-existing chronic diseases include diabetes, asthma, cardiovascular diseases
Table 2. Residential characteristics of the study participants
| Variable |
|----------|
| **Categorical variables (n=149)** | n (%) |
| Neighborhood greenspace | |
| Access to small neighborhood greenspace | |
| No | 9 (6.0) |
| Yes | 140 (94.0) |
| Access to large neighborhood greenspace | |
| No | 36 (24.2) |
| Yes | 113 (75.8) |
| Eurostat urbanization | |
| Cities | 56 (37.6) |
| Towns and suburbs | 93 (62.4) |
| Rural | 0 |
| Noise levels | |
| L_{den} < 53 dB | 97 (65.1) |
| L_{den} ≥ 53 dB | 47 (31.5) |
| Missing | 5 (3.4) |
| **Continuous variables** | **Geometric mean (95% CI)** |
| Distance to major roads (m) (n=149) | 290 (240, 349) |
| 2\textsuperscript{nd} trimester air pollution (µg/m\textsuperscript{3}) (n=133) | |
| NO\textsubscript{2} - 90 days prior to sampling | 23.03 (21.67, 24.47) |
| NO\textsubscript{2} - 1 year prior to sampling | 24.55 (23.39, 25.76) |
| PM\textsubscript{2.5} - 90 days prior to sampling | 11.61 (11.06, 12.21) |
| PM\textsubscript{2.5} - 1 year prior to sampling | 13.09 (12.81, 13.37) |
| BC - 90 days prior to sampling | 1.13 (1.05, 1.21) |
| BC - 1 year prior to sampling | 1.29 (1.24, 1.36) |
| 3\textsuperscript{rd} trimester air pollution (µg/m\textsuperscript{3}) (n=78) | |
| NO\textsubscript{2} - 90 days prior to sampling | 23.19 (21.53, 24.98) |
| NO\textsubscript{2} - 1 year prior to sampling | 24.70 (23.27, 26.23) |
| PM\textsubscript{2.5} - 90 days prior to sampling | 11.55 (10.95, 12.18) |
Note: Categorical data is described as frequencies (%), continuous data is described by geometric mean with 95% confidence interval (95% CI). All data is based on the maternal residential address. Major roads include E- or N-roads. Access to small neighborhood greenspace is defined as access to > 0.2 hectares (ha) of greenspace within a travel distance of 400 meters (m) from residence, access to large neighborhood greenspace is defined as access to > 10 ha of greenspace within a travel distance of 800 m from residence. \( L_{\text{den}} \), day–evening–night noise level; NO\(_2\), nitrogen dioxide; PM\(_{2.5}\), fine particulate matter with an aerodynamic diameter \( \leq \) 2.5 \( \mu \)m; BC, black carbon.

Table 3. Spearman rank correlations between residential exposures in the 2nd trimester study population (n=133)

|       | BC   | NO\(_2\) | Distance to major road | Small greenspace | Large greenspace | Noise     | ADI      |
|-------|------|----------|------------------------|------------------|------------------|-----------|----------|
| PM\(_{2.5}\) | 0.68* | 0.68*    | -0.15                  | -0.02            | -0.04            | 0.08      | 0.15     |
| BC    | 0.89* |          |                        |                  |                  |           |          |
| NO\(_2\) |      | -0.37*   |                        |                  |                  |           |          |
| Distance to major road | 0.01 |          |                        |                  |                  |           |          |
| Small greenspace |      | 0.43*    |                        |                  |                  |           |          |
| Large greenspace |      | -0.20*   |                        |                  |                  |           |          |
| Noise  |      |          |                        |                  |                  | 0.21      |          |

*Significant correlations (p < 0.05)

Note: Air pollutants were modelled at the maternal home address, 3-mont mean concentrations were calculated. Major roads include E- or N-roads. Access to small neighborhood greenspace is defined as access to > 0.2 hectares (ha) of greenspace within a travel distance of 400 meters (m) from residence, access to large neighborhood greenspace is defined as access to > 10 ha of greenspace within a travel distance of 800 m from residence. \( L_{\text{den}} \), day–evening–night noise level. ADI, area deprivation index (%).
deprivation index; NO₂, nitrogen dioxide; PM₂.₅, fine particulate matter with an aerodynamic diameter ≤ 2.5 μm; BC, black carbon.

Table 4. Spearman rank correlations between residential exposure in the 3rd trimester study population (n=81)

|          | BC  | NO₂  | Distance to major road | Small greenspace | Large greenspace | Noise | ADI |
|----------|-----|------|------------------------|------------------|------------------|-------|-----|
| PM₂.₅   | 0.61*| 0.59*| -0.06                  | -0.19*           | -0.31*           | 0.01  | 0.04|
| BC      | 0.89*| -0.24*| -0.22*                 | -0.3*            | 0.18             | 0.31* |
| NO₂     | -0.32*| -0.21*| -0.28*                 | 0.24*            | 0.39*            |
| Distance to major road | 0.13 | 0.26* | -0.31*                 | -0.25*           |
| Small greenspace |       | 0.41* | -0.37*                 | -0.08            |
| Large greenspace |     |       |                        |                  |
| Noise   |     |       |                        |                  | 0.25*            |

*Significant correlations (p < 0.05)

Note: Air pollutants were modelled at the maternal home address, 3-mont mean concentrations were calculated. Major roads include E- or N-roads. Access to small neighborhood greenspace is defined as access to > 0.2 hectares (ha) of greenspace within a travel distance of 400 meters (m) from residence, access to large neighborhood greenspace is defined as access to > 10 ha of greenspace within a travel distance of 800 m from residence. Noise exposure is evaluated as exposure above the WHO health-based guideline of 53 dB L_{den} (day–evening–night noise level). ADI, area deprivation index; NO₂, nitrogen dioxide; PM₂.₅, fine particulate matter with an aerodynamic diameter ≤ 2.5 μm; BC, black carbon.

Table 5. Associations between residential exposures and hair cortisol concentrations in the second and third pregnancy trimester
| Exposure               | 2nd trimester unadjusted model (n=133) | 2nd trim HCC adjusted model a (n=133) | 3rd trimester unadjusted model (n=81) |
|-----------------------|---------------------------------------|--------------------------------------|--------------------------------------|
|                       | p-value | β (95% CI) | p-value | β (95% CI) | p-value | β (95% CI) |
| 3-month mean PM$_{2.5}$ | 0.009   | 0.81 (0.70, 0.95) | 0.200   | 0.87 (0.70, 1.08) | 0.227* | 1.20 (0.89, 1.62) |
| 3-month mean NO$_2$    | 0.934   | 0.99 (0.83, 1.18) | 0.287   | 1.10 (0.92, 1.34) | 0.016* | 1.42 (1.07, 1.88) |
| 3-month mean BC        | 0.551   | 0.94 (0.79, 1.13) | 0.775   | 1.03 (0.84, 1.27) | 0.032* | 1.37 (1.03, 1.82) |
| Dist. to major roads   | 0.016   | 0.82 (0.70, 0.96) | 0.011   | 0.82 (0.70, 0.95) | 0.040   | 0.74 (0.55, 0.99) |
| Small greenspace       | 0.117   | 0.68 (0.42, 1.10) | 0.061   | 0.63 (0.38, 1.02) | 0.354   | 0.67 (0.28, 1.05) |
| Large greenspace       | 0.073   | 0.77 (0.57, 1.03) | 0.095   | 0.78 (0.59, 1.04) | 0.062   | 0.65 (0.41, 1.02) |

Note: aadjusted for season of sampling. Estimates are presented for an increase of exposure from the 25th to the 75th percentile. *Data available of 78 participants. Air pollutants were modelled at the maternal home address, 3-mont mean concentrations were calculated. Major roads include E- or N-roads. Access to small neighborhood greenspace is defined as access to > 0.2 hectares (ha) of greenspace within a travel distance of 400 meters (m) from residence, access to large neighborhood greenspace is defined as access to > 10 ha of greenspace within a travel distance of 800 m from residence. CI, confidence interval; NO$_2$, nitrogen dioxide; PM$_{2.5}$, fine particulate matter with an aerodynamic diameter ≤ 2.5 μm; BC, black carbon

Figures
Figure 1

Interaction between distance to a major road and access to a large neighborhood greenspace in relation to 2nd trimester HCC
Figure 2

Interaction between distance to a major road and access to a large neighborhood greenspace in relation to 3rd trimester HCC

Supplementary Files

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- supplement6.docx