A review on the application of the exposome paradigm to unveil the environmental determinants of age-related diseases

Enmin Ding¹,², Yu Wang¹, Juan Liu¹, Song Tang¹,³ and Xiaoming Shi¹,³*

Abstract
Age-related diseases account for almost half of all diseases among adults worldwide, and their incidence is substantially affected by the exposome, which is the sum of all exogenous and endogenous environmental exposures and the human body’s response to these exposures throughout the entire lifespan. Herein, we perform a comprehensive review of the epidemiological literature to determine the key elements of the exposome that affect the development of age-related diseases and the roles of aging hallmarks in this process. We find that most exposure assessments in previous aging studies have used a reductionist approach, whereby the effect of only a single environmental factor or a specific class of environmental factors on the development of age-related diseases has been examined. As such, there is a lack of a holistic and unbiased understanding of the effect of multiple environmental factors on the development of age-related diseases. To address this, we propose several research strategies based on an exposomic framework that could advance our understanding—in particular, from a mechanistic perspective—of how environmental factors affect the development of age-related diseases. We discuss the statistical methods and other methods that have been used in exposome-wide association studies, with a particular focus on multiomics technologies. We also address future challenges and opportunities in the realm of multidisciplinary approaches and genome–exposome epidemiology. Furthermore, we provide perspectives on precise public health services for vulnerable populations, public communications, the integration of risk exposure information, and the bench-to-bedside translation of research on age-related diseases.

Keywords: Exposomic framework, Aging, Statistical approach, Multiomics, Precise public health

Background
Due to the continuous advancement of biomedical technologies and the substantial improvement in living conditions worldwide, human lifespans have increased dramatically over the past century [1]. Nevertheless, the prevalence of age-related diseases remains considerable: 92 (31.4%) of the 293 diseases listed in the Global Burden of Diseases, Injuries, and Risk Factors Study (GBD) 2017 were identified as age-related, and these accounted for 51.3% of the worldwide burden of disease in adults [2]. In addition, age-related diseases account for two-thirds of human deaths globally and 90% of all deaths in industrialized nations [3]. Moreover, sociodemographic index (SDI) analysis revealed that the rates of age-related diseases ranged from 137.8 disability-adjusted life years (DALYs) per 1000 adults in high-SDI countries to 265.9 DALYs per 1,000 adults in low-SDI countries [2]. The most common age-related diseases are neurological degenerative diseases (e.g., Parkinson's disease [PD] and Alzheimer’s disease [AD] or dementia), chronic
obstructive pulmonary disease (COPD), coronary artery disease (CAD), stroke, type 2 diabetes mellitus (T2DM), and senile deafness [2, 4]. These diseases place an enormous economic and psychological burden on patients, their families, and societies worldwide [5].

In 2021, Science published a special issue entitled “125 Questions: Exploration and Discovery.” One of these 125 questions was “Can we stop ourselves from aging?” The U.S. National Institute on Aging (NIA) at the National Institutes of Health (NIH) states that “aging is associated with changes in dynamic biological, physiological, environmental, psychological, behavioral, and social processes.” Although geneticists and epidemiologists have long debated the relative importance of the role played by genotype or the environment in the development of age-related diseases, it is apparent that both can play substantial roles in this process [6, 7]. However, most etiological studies have concentrated on the role of genotype and have considered the environment to play a secondary role. Nevertheless, an analysis of GBD data showed that nearly 50% of deaths worldwide are attributable to environmental exposure, primarily exposure to airborne particulates (including household air pollution and occupational exposure; 14% of all deaths), smoking and secondhand smoke (13%), plasma sodium concentrations (6%), and alcohol consumption (5%) [8]. In contrast, a recent analysis of 28 chronic diseases in identical twins showed that the genetic-related risks of developing one of five age-related diseases were 33.3%, 10.6%, 36.3%, 19.5%, and 33.9% for AD, PD, CAD, COPD, and T2DM, respectively, with a mean of only 26% [9]. The results of over 400 genome-wide association studies (GWASs) have also elucidated that the heritability of degenerative diseases is only approximately 10% [10, 11]. Consequently, non-genetic drivers, such as environmental factors, are now recognized as major risk factors for age-related diseases. The contributions of environmental factors to the development of age-related diseases can be revealed by analyses of all of the factors to which individuals are exposed in their life and the relationships between these exposures and age-related diseases [12, 13].

The concept of the “exposome” was established by C.P. Wild in 2005 as an environmental analogy of the genome and represents the sum of all exposures of an individual throughout the life course [12]. Wild described three categories of the exposome that should be evaluated: (a) specific external exposures, (b) general external exposures, and (c) internal exposures [14]. In this review, we describe the environmental drivers of typical age-related chronic diseases, mainly focusing on AD, PD, CAD, COPD, T2DM, and senile deafness within an exposomic framework. We propose representative strategies for exposomic research on age-related diseases and describe the measurements and statistical approaches that could be used. In addition, we note the challenges of exposomic research and recommend guidelines for the implementation of exposomic concepts in research on age-related diseases.

Epidemiologic studies on the environmental causes of age-related diseases

Specific external exposures

Air pollution

Atmospheric air pollution is the leading form of specific external exposures, and its role in the development of age-related diseases has been extensively studied. For example, a cohort study in Ontario, Canada, evaluated the association between air pollution and incident PD, which showed that ambient 2.5-µm particulate matter (PM2.5) was significantly associated with a 4% increase in incident PD (95% confidence interval [CI] 1.01, 1.08) [15]. Similarly, high concentrations of nitrogen oxides have been found to be significantly associated with an increased risk of AD or dementia in cohort studies in Sweden (hazard ratio [HR]: 1.38; 95% CI 0.87, 2.19) and Taiwan (HR: 1.54; 95% CI 1.34, 1.77) [16, 17]. In the past decade, many studies have explored the relationships between air pollutant exposure and COPD and have determined that the main airborne pollutants affecting the respiratory system are PM2.5, 10-µm particulate matter (PM10), ozone, carbon monoxide (CO), and sulfur dioxide (SO2) [18, 19]. Moreover, a systematic review and meta-analysis showed that short-term exposure to CO (relative risk [RR]: 1.015; 95% CI 1.004, 1.026), SO2 (RR: 1.019; 95% CI 1.011, 1.027), nitrogen dioxide (NO2) (RR: 1.014; 95% CI 1.009, 1.019), PM2.5 (RR: 1.011; 95% CI 1.011, 1.012), and PM10 (RR: 1.002; 95% CI 1.000, 1.004) was significantly associated with hospitalization or death due to stroke [20]. Furthermore, multiple cohort studies have shown that exposure to high concentrations of NO2 and PM2.5 was associated with an increased risk of T2DM [21–25].

Chemical contaminants

Some toxic chemicals (e.g., pesticides and heavy metals) have been confirmed to be significantly associated with higher risks of age-related diseases. A meta-analysis of 10 cohort and case–control studies identified that 5 and 10 years of pesticide exposure were associated with a 5% (95% CI 1.02, 1.09) and 11% (95% CI 1.05–1.18) increase in the risk of PD, respectively [26]. A cohort study of 1500 older adult participants aged 65 years or older in southwest France found that occupational exposure to pesticides (e.g., fungicides) was significantly associated with AD in men (RR: 2.29; 95% CI 1.02, 5.63) [27]. Numerous other studies—such as the Agricultural Health Study, the
increased the incidence of coronary heart disease and vian population study concluded that road traffic noise.

Similarly, a recent meta-analysis and a large Scandinavian study investigating that noise aggravates the symptoms of dementia [37]. Due to dementia (RR: 1.15; 95% CI 1.11, 1.20), indicating that noise increases the risk of AD in men (RR: 2.30; 95% CI 1.00, 5.30) [36]. Moreover, short-term exposure to traffic noise was found to be associated with hospitalization due to dementia (RR: 1.15; 95% CI 1.11, 1.20), indicating that noise aggravates the symptoms of dementia [37].

Electromagnetic fields and noise are the most common risk factors for age-related diseases, such as neurodegenerative, cardiovascular, and cerebrovascular diseases. A follow-up of 931 participants aged 75 years and older without dementia in Sweden found that high-term occupational exposure to a high electromagnetic field might increase the risk of AD in men (RR: 3.35; 95% CI 1.49, 7.52) [33]. However, there have been fewer studies of associations of age-related diseases with internal exposures to emerging organic contaminants, such as perfluoroalkyl substances and brominated flame retardants. In one example, the Nurses’ Health Study II determined that higher plasma concentrations of perfluorooctanesulfonic acid (PFOS) or perfluorooctanoic acid (PFOA) was associated with an increased risk of T2DM [28–30].

Lifestyle and diet
The lifestyle factors most significantly associated with the risk of age-related diseases are smoking, passive smoking, and drinking alcohol. The European Prospective Investigation into Cancer and Nutrition for Parkinson’s Disease, a prospective European population-based cohort study of 220,494 people aged 37 to 70 from 13 centers in 8 countries, found a causal relationship between smoking and PD (HR: 0.70; 95% CI 0.49, 0.99) [40]. A review summarized the existing evidence and confirmed smoking as the main risk factor for COPD [41]. Other environmental exposures, such as dietary exposures and secondhand smoke exposure during pregnancy or early childhood, may also be important risk factors for COPD. A recent cross-sectional analysis of 164,770 adults aged 40 to 69 from the U.K. Biobank found that smoking and passive smoking were associated with a risk of senile deafness (ORs: 1.15 [95% CI 1.09, 1.21] and 1.28 [95% CI 1.21, 1.35], respectively) [42]. Analogously, a recent study of 1,787 mid-to-late-aged adult participants concluded that added salt in the diet might increase at-risk individuals’ risk of AD, while consuming cheese and red wine on a daily basis and lamb on a weekly basis might improve long-term cognitive outcomes [43]. Patel et al. conducted the first exposome-wide association study (EWAS) on T2DM using data from National Health and Nutrition Examination Survey (NHANES) cohorts from 1999 to 2006 [44]. Logistic regression models with multiple comparisons were used, and significant associations for vitamin γ-tocopherol (OR: 1.5; 95% CI 1.3, 1.7) and β-carotenes (OR: 0.6; 95% CI 0.5, 0.7) were associated with T2DM.

Physical factors
Electromagnetic fields and noise are the most common risk factors for age-related diseases, such as neurodegenerative, cardiovascular, and cerebrovascular diseases. A follow-up of 931 participants aged 75 years and older without dementia in Sweden found that long-term occupational exposure to a high electromagnetic field might increase the risk of AD in men (RR: 2.30; 95% CI 1.00, 5.30) [36]. Moreover, short-term exposure to traffic noise was found to be associated with hospitalization due to dementia (RR: 1.15; 95% CI 1.11, 1.20), indicating that noise aggravates the symptoms of dementia [37]. Similarly, a recent meta-analysis and a large Scandinavian population study concluded that road traffic noise increased the incidence of coronary heart disease and stroke by 8% (95% CI 1.01, 1.15) and 6% (95% CI 1.03, 1.08) per 10 A-weighted decibels [dB(A)], respectively [38, 39].

Internal exposures
Medication
The use of antipsychotics, hormones, and anticholinergics has been found to significantly increase the risks of PD and AD. For example, it has been found that among older adults, the use of antipsychotics, such as phenothiazine (RR: 3.65; 95% CI 1.41, 9.45) or benzamide (RR: 2.59; 95% CI 1.23, 5.43), and particularly trichloroethylene (OR: 6.1, 95% CI 1.2, 33), may increase the risk of PD [45, 46]. Moreover, a prospective population-based cohort study in Seattle, Washington (USA), found that higher cumulative anticholinergic use was significantly associated with an increased risk of dementia (HR: 1.54; 95% CI 1.21, 1.96) [47].
**Gut microbiota**
Gut microbial dysbiosis has been found to be associated with many age-related diseases. For example, the fecal samples of PD patients were found to have a substantially lower abundance of *Prevotellaceae* than those of healthy individuals. Additionally, a direct correlation was found between the fecal abundance of *Enterobacteriaceae* and the severity of postural instabilities and gait difficulties [48]. Moreover, the fecal samples of 64 Italian patients with PD exhibited low abundances of gut microbiota (e.g., *Lachnospiraceae*) linked to anti-inflammatory/nerveprotective effects and abnormal concentrations of several classes of fecal metabolites (lipsids, vitamins, amino acids, and other organic compounds) [49]. Vogt et al. characterized the bacterial taxonomic composition of fecal samples from participants with and without a diagnosis of dementia due to AD and revealed that the gut microbiome of AD participants had less microbial diversity and was compositionally distinct from that of control individuals. They also identified phylum- to genus-wide differences in bacterial abundance in the microbiome of AD participants, such as a decreased abundance of *Firmicutes*, an increased abundance of *Bacteroidetes*, and a decreased abundance of *Bifidobacterium* [50].

**Inflammation and oxidative stress**
Inflammation is the basis of aging and many age-related chronic diseases, which in turn increase the rate of aging. Walker et al. examined the association between systemic inflammation measured during midlife and 20-year cognitive decline within the Atherosclerosis Risk in Communities cohort study. An increase in the midlife inflammation composite score was associated with an additional 20-year decline of $-0.035$ SDs (95% CI $-0.062$, $-0.007$) in the cognitive composite score [51]. Gong et al. followed three cohort studies in the USA and explored whether proinflammatory diets were associated with increased CVD risks. During 5,291,518 person-years of follow-up, higher dietary inflammatory pattern scores were predefined based on levels of 3 systemic inflammatory biomarkers, which were associated with an increased risk of CVD (HR: 1.38; 95% CI 1.31, 1.46) [52]. Accordingly, age-related diseases can be partly conceptualized as manifestations of accelerated inflammation or aging [53].

**Hormones**
Hormone exposure has been associated with the risk of many age-related diseases, especially among postmenopausal women. Savolainen-Peltonen et al. investigated hormone exposure and AD among Finnish postmenopausal women using Finnish national population and drug register data (1999–2013). The results indicated that long-term use of hormone therapy might be associated with an increased risk of AD (OR: 1.17, 95% CI 1.13, 1.21) [54]. In addition, the pooled results of a systematic review and a time–response meta-analysis showed that there was a significant association between hormone therapy and the risk of AD (OR: 1.08; 95% CI 1.03, 1.14) in menopausal women. However, no comparable association was uncovered for PD [55].

**General external exposures**

**Green space and urbanization**
Less adequate evidence has been found for an association between green spaces and cognitive function. In the Cognitive Function and Ageing Studies in England, a higher risk of dementia was found among people who were in the highest quartile of neighborhood natural environment availability than among those in lower quartiles of neighborhood natural environment availability [56]. In contrast, a longitudinal study in the UK demonstrated that a higher level of surrounding greenness in residential areas was associated with a slower cognitive decline among residents over a 10-year follow-up (0.020; 95% CI 0.003–0.037 per interquartile increment) and that this was independent of air pollution exposure [57]. Similarly, a comparison of the effects of the indicators of the built environment on the risk of disease found that urbanization was associated with a higher incidence of COPD (1.05; 95% CI 1.01–1.08 per interquartile increment), while residential green space was associated with a lower incidence of COPD (0.89; 95% CI 0.84–0.93 per interquartile increment) [58].

**Socioeconomic status and educational level**
In a prospective cohort study, social integration was found to contribute to a lower risk of cardiovascular disease (HR: 0.67; 95% CI 0.53, 0.86). In addition, people with the highest social integration level had reduced mortality risks of 30%, 47%, and 53% for AD, COPD, and T2DM, respectively, compared with those with the lowest social integration level [59]. A systematic review and meta-analysis of long-term studies determined that a low social participation index was related to the risk of dementia in terms of social support (RR: 1.28; 95% CI 1.01, 1.62) and social networking (RR: 1.59; 95% CI 1.31, 1.96), while a high social participation index had a moderately protective effect (RR: 0.88; 95% CI 0.80, 0.96) [60]. A cross-sectional study of adults aged 60 or older in China indicated that fewer years of education (OR: 1.55; 95% CI 1.38, 1.73) and being widowed, divorced, or living alone (OR: 2.66; 95% CI 2.29, 3.10) were risk factors for dementia [61]. Another cross-sectional study of older adults with a mean age of 75 years in Denmark showed that compared with individuals with lower household
incomes, those with a higher household income were less likely to receive a dementia diagnosis after referral (OR: 0.65; 95% CI 0.55, 0.78) [62]. The search strategies and main results of epidemiologic studies reporting the association between environmental risk factors and age-related diseases (Table S1) in this review are shown in the Additional file 1.

Potential hallmarks linking the exposome to age-related diseases

Age-related diseases are thought to arise from common underlying processes that precipitate molecular changes over time [63]. These shared mechanisms are grouped into the following hallmarks of aging: genomic instability, epigenetic alterations, telomeric attrition, loss of proteostasis, deregulated nutrient sensing, mitochondrial dysfunction, oxidative stress/damage, inflammation, cellular senescence, stem cell exhaustion, and altered intercellular communication [64, 65]. Some studies have reported the effects of environmental exposures on various hallmarks of aging. For example, exposure to heavy metals and organic pollutants has been associated with mitochondrial dysfunction and DNA methylation alteration in AD [66, 67], and early embryonic exposure to 2,3,7,8-tetrachlorodibenzo-p-dioxin may alter global DNA methylation patterns, increasing the risk of T2DM [68]. Bisphenol A concentrations were also found to be associated with increased senescence, inflammation, and decreased telomere lengths in patients with T2DM [69]. However, there have been few other consistent findings confirming that aging hallmarks link the exposome to risks of age-related diseases.

In particular, the alteration of gene expression through DNA methylation, histone modification, and noncoding RNA-associated gene silencing are primary targets of environmental insults considered to induce and sustain epigenetic changes that might result in age-related diseases [70]. In that regard, advances in technology have made it feasible to assay methylation status across the genome in a robust, high-throughput manner, allowing us to address this issue in a data-driven manner. To date, numerous studies have detailed the pattern by which global or site-specific DNA methylation is affected by environmental triggers [67]. For instance, Prado-Bert et al. analyzed the association between the early-life exposome and epigenetic age acceleration based on the Human Early-Life Exposome (HELIX) project. Indoor particulate matter absorbance (PMabs) and parental smoking were found to be positively associated with age acceleration, which was calculated based on Horvath’s skin and blood clock for 1173 children [71]. However, relatively few researchers are looking to understand the functional consequences of methylation marks by assessing gene, protein, and metabolite expression. Recently, Everson et al. identified 443 CpGs that were associated with maternal smoking exposure during pregnancy in seven American, Australian, and European studies. Subsequent expression quantitative trait methylation analyses testing the associations between maternal smoking exposure-associated CpGs and the expression of nearby mRNA showed that exposure-associated CpGs were enriched for environmental response, growth factor signaling, and inflammation [72].

Limitations of studies examining whether environmental factors contribute to the development of age-related diseases

Although the link between environmental exposure and age-related diseases has been extensively investigated, there are three limitations to previous studies in this area.

1. Most exposure assessments have examined only one chemical or class of chemicals via a traditional targeted measurement approach. However, real-world exposures do not occur in such a manner, and thus, “broad-spectrum” exposure assessments are needed, together with systematic screening of key exposomic elements as well as optimized consideration of the components of exposome domains that affect age-related diseases [73].

2. Exposure–health association studies have not fully considered the dynamic changes in exposure or the effects of (mixed) exposures in early life stages or in other sensitive time windows on the risks of age-related diseases.

3. Studies have rarely considered chemical absorption, distribution, metabolism, and excretion processes or endogenous biomolecular responses, such as those involving hormones, metabolites, infections, inflammatory reactions, fat peroxidation, oxidative stress, and aging hallmarks.

Therefore, the link between environmental exposure and age-related diseases needs to be explored within an exposomic framework. Such an approach would systematically and substantially increase our understanding of the environmental drivers of age-related diseases, reveal how environmental exposure interacts with genetic susceptibility to age-related diseases, and elucidate the potential biological mechanisms of how these factors affect disease susceptibility, development, and progression.

Strategies for exposomic research of age-related diseases

Given the limitations of recent research and the evolution of the exposome concept, we propose the following three major strategies for exposomic research of age-related diseases.
(1) Comprehensive determination of the chemical compositions of biological fluids and standardized assessment. Pioneers in exposomic research, such as Rappaport, Smith, and Miller, have emphasized the need to fully characterize the chemical composition of biological fluids (e.g., blood and urine) for precisely assessing individuals' immediate environments and thus their risk of age-related diseases [74, 75]. In addition, standardization of the exposure/outcome assessment methods using conventional tools along with innovative technologies as well as standardized quality assessment tools may contribute to the utility of the exposome concept [73].

(2) Consideration of the roles of biological responses and endogenous processes, especially aging biomarkers, at various omics levels (i.e., at the level of the epigenome (e.g., DNA methylation clock), the transcriptome, the proteome, the metabolome, and the microbiome) and with respect to extracellular RNA (including microRNA and long non-coding RNA) abundance and telomere length [76]. Alterations in biomarkers compromise cell and tissue functions and contribute to the incidence of age-related diseases, leading to loss of function and death [77]. In addition, the concentrations of aging biomarkers can reflect the physiological state of individuals and the underlying molecular mechanisms (e.g., genomic instability, epigenetic perturbation, telomeric attrition, loss of proteostasis, altered nutrient sensing, mitochondrial dysfunction, cellular senescence, and disrupted intercellular communications) related to the exposome throughout individuals' lifespans [64, 78].

(3) Performance of exposure and outcome measurements in critical time windows during the aging process. The exposome includes a time dimension, as it emphasizes the importance of assessing exposure throughout life, beginning at the moment of conception. However, a lifelong study of all environmental factors would be highly challenging to perform in real life [79]. Moreover, people age at dissimilar rates, even in terms of tissues and organs, which have a tissue-specific aging signature [80]. In addition, environmental exposure to harmful factors in different critical life windows of susceptibility may lead to distinct disease phenotypes [81, 82]. However, a recent study indicated that there appear to be distinct “waves” of aging over the course of a lifetime at the population level at three ages: 34, 60, and 78 years old [83]. This suggests that exposomic measurements in critical time windows may be useful for the study of the risk of age-related diseases.

In summary, systematic approaches are needed to integrate diverse information in exposome research (Fig. 1). The comprehensive evaluation of the relationships between exposures and associated biological responses, such as epigenetic modifications and metabolites, would yield valuable insights into the key elements of the exposome and how environmental factors and their interactions contribute to the risk and development of age-related diseases.

### Exposomic measurement tools for use in age-related disease studies

Exposomes can be measured at the population level using geographic information systems (GISs) and remote sensing technologies (such as satellites) or at the individual level using records and surveys, personal sensors, and biological samples. The exposomic information of environmental and biological samples can be measured using advanced analytical techniques, such as high-resolution mass spectrometry (HRMS) and omics platforms.

#### Population-level exposures

**GISs**

The use of GISs has changed environmental health research, as GISs integrate databases that connect various attribute data based on geographic location (e.g., residential address). Therefore, GISs can quantify the buffer distance between exposure sources and human receptors and can be used to describe the proximity of roads, factories, green spaces, water bodies, and other features that may result in exposures that negatively (e.g., farms, which may result in exposure to pesticides) [84] or positively (e.g., health food stores or entertainment venues) affect human health [85].

**Remote sensing**

Remote sensing is the science of obtaining information about objects or regions from a distance (usually from airplanes or satellites). Remote sensing technology has been used to estimate the levels of various environmental exposures, such as PM<sub>2.5</sub> [86], NO<sub>2</sub> [87], green space [88], temperature [89], building environment [90], outdoor light at night [91], and pesticides [92]. For example, global PM<sub>2.5</sub> concentrations were estimated from information from three satellites in conjunction with a chemical transport model and ground-based sun photometer observations [93]. Remote sensing estimates have also been used to assess the relationships between PM<sub>2.5</sub> exposure and cardiovascular diseases [25, 94, 95]. Moreover, recent studies have used a 1-km valuation of U.S. PM<sub>2.5</sub> concentrations, thereby increasing the utility of such measurements in exposomic studies [96, 97].
Individual-level exposures

**Portable/personal sampling and sensing devices**

An increasing number of personal sensors are being developed to monitor heart rate, blood sugar concentrations, blood pressure, muscle activity, temperature, and sweat production [98]. Remarkable advances in novel algorithms in wearable technologies are allowing for accurate acquisition of personal temperature exposures, chemical and biological exposures, and clinical laboratory measurements to explore individual-level 24-h exposures in various microenvironments in exposome studies [99–102]. Koelmel et al. captured the changes in gas- and particulate-phase airborne chemicals to which individuals were exposed and determined which exposures posed the greatest risk to health. They used wearable exposure monitors, i.e., passive air-sampling wristbands, which were provided to 84 healthy participants (aged 60–69) as part of the Biomarkers for Air Pollution Exposure (China BAPE) study [103, 104]. Jiang et al. developed a sensitive method to monitor personal airborne biological and chemical exposures, which enabled them to follow the personal exposomes of 15 individuals for up to 890 days over 66 distinct geographical locations. They found that humans are exposed to thousands of microbial species with great intraspecies diversity and demonstrated that the human exposome is highly dynamic and influenced by spatial/lifestyle and seasonal variables [105, 106].

**Records and surveys**

Data from health records and surveys can be used to describe the exposure characteristics of individuals at a relatively low cost. For example, electronic health records provide an efficient way to estimate individual-level exposure to, for example, pharmaceutical drugs, alcohol, or cigarette smoke. Similarly, surveys performed as part of epidemiological studies can provide valuable information on peoples’ diet, socioeconomic status, education level, and psychosocial stress level.

**Biological samples**

The analysis of biological samples (e.g., blood, urine, saliva, feces, and hair) enables quantification of the burden of chemical exposures for an individual. In particular, abundant information on the biological response to an exposure can be obtained by quantifying small molecules in biological samples, such as inflammatory factors and metabolites.
Analytical platforms

**HRMS**

Advancements in HRMS have provided new tools to identify and quantify multiple chemical substances in a wide range of media and have extended the analytical window beyond the targeted profiling of known metabolites and priority pollutants [107]. In particular, HRMS enables the simultaneous analysis of many exogenous and endogenous compounds and the characterization of the state of a biological system and its response to environmental factors [13, 108].

Nontargeted HRMS-based analysis can collect data on thousands of chemical features in a single analytical run and is thus suitable for mass screening, although most of these features remain unannotated [109]. However, targeted HRMS-based analysis involves performing multiple analyses of known and specific chemicals, including absolute qualitative and quantitative analyses of the substances to be measured using certified standards, and is therefore suitable for validation in exposomic studies. For example, as part of the China BAPE study [104], Guo et al. performed targeted HRMS-based analysis of 70 airborne compound exposures of 84 elderly people in Jinan, China, and thereby described their personal exposure to air pollutants in terms of concentration and distribution [110].

**Multiomics technology**

The broad diversity of omics biomarkers that have been used to assess biological responses provides new opportunities to understand the impact of the environment on the risk of age-related diseases. For example, the multiomics analysis and integration method produces a priority list of multiple sets of biomarkers, which together reflect the molecular responses of the exposome. Each of these data warrants integration into a biomarker panel to aid physicians in developing age-related disease diagnoses and prognoses [78].

By integrating the response measures of genomics, epigenomics, transcriptomics, proteomics, and metabolomics, it is possible to develop a systems biology-level understanding of how exposure affects key biochemical processes. Similarly, the aggregated biological response model, which combines toxicology and pharmacology with molecular and environmental epidemiology, represents a new paradigm for delineating the mechanisms of chemical toxicology [13]. The further development of statistical methods to determine the interactions between biological response networks [111, 112] and the application of multiomics methods in cohort studies to elucidate the characteristics of human exposure [113, 114] will facilitate exposomic discoveries in the years ahead. For instance, the EXPOSOMICS project, an integration of 14 different studies, applied multiomics technologies such as metabolomics and adductomics to quantify exposure levels and to measure downstream effects of environmental exposures by the epigenome, transcriptome, and proteome [114]. The PISCINA-II study (part of the EXPOSOMICS project) included 60 volunteers who swam for 40 min in a chlorinated pool, and metabolome-wide association showed that disinfectant byproducts were associated with 333 metabolic features, of which 13 metabolites were identified and enriched in the tryptophan metabolism pathway [115].

Recently, more research has incorporated multiomics technologies and personal sampling and sensing devices to reveal the dynamic molecular effects of precise personal exposure. For example, Jiang et al. developed a personal wearable device that monitored airborne biological and chemical exposures and followed the personal exposomes of 15 individuals for nearly 3 years [105]. Next-generation sequencing and liquid chromatography–mass spectrometry (LC–MS) technologies were then used to detect the biotic/chemical exposome and found that the human exposome was dynamic and influenced by various environmental and spatial/lifestyle variables. Gao et al. examined the effects of environmental exposures on well-dissected personal internal health changes based on longitudinal individual exposomes and internal multiomics (gut microbiome, proteome and metabolome) [101]. By investigating the correlations of exposomics and the exposome and clinical data, it was found that changes in external exposures could be associated with thousands of biomolecular changes in the body involved in immune, kidney, and liver functions.

**Evolving statistical methods in exposome–health association studies of age-related diseases**

**Exposome-wide association study (EWAS)**

An EWAS is a statistical method similar to a GWAS that is used to continuously and independently detect the association of many exposures with a set of results (with adjustment for potential confounding factors). Although nonideal, for practical purposes—because exposure studies examine many exposures—the same set of confounding factors is typically used for all exposures. Furthermore, multiple comparison corrections should be applied to minimize the number of false-positive results obtained. Bonferroni–Holm correction method (family-wise error rate [FWER] method) or the Benjamini–Hochberg correction method (false discovery rate [FDR] method) are among the most commonly used multiple comparison correction methods in epidemiological research. However, a limitation is that these methods assume that tests are independent, and if this assumption is violated, these methods may generate highly
conservative results [116]. The Bonferroni correction method has been devised and involves dividing the significance level $\alpha$ by the test number $M$ ($\alpha_{\text{corrected}} = \frac{\alpha}{M}$); this changes the value of $M$ according to the number of valid checks, as determined by the relevant structure of the data [116, 117]. Methods such as stepwise multiple testing have been developed to capture the joint dependency structure of the test statistic, thereby improving the ability to detect false hypotheses [118]. However, there has been no development of multiple-detection correction methods to simultaneously deal with the diversity of exposure and the results of exposure environments. In a study that simulated the true correlation structure of the Infancia y Medio Ambiente (INMA, Spain) birth cohort exposure variables, EWAS exhibited a higher FDR than other methods but showed the greatest sensitivity [119].

**Variable selection tools**

The most popular variable selection techniques in association research include the deletion/substitution/addition (DSA) algorithm, the elastic net (ENET) algorithm, and the graphical unit evolutionary stochastic search (GUESS) algorithm. For example, the DSA algorithm was used in the recent HELIX project to explore the relationship between early-life exposure to environmental hazards and childhood lung function and the relationship between urban pregnancy exposure to environmental hazards and birth weight [120, 121]. In simulation research, the DSA method has shown higher sensitivity and a lower FDR than other methods and good performance in capturing interactive items [119, 122]. However, the DSA method has routinely been criticized because its estimations are inconsistent if the ratio of the sample size to the number of candidate predictions is small, and its confidence intervals are low if there is a large correlation between predictions [123].

The ENET algorithm has been used to study the relationship between various environmental pollutants and birth weight in three cohorts, one each in Greenland, Poland, and Ukraine, and to examine the relationship between concentrations of several persistent organic pollutants in breast milk samples and infant behavioral problems [124, 125]. In addition, the ENET method showed high sensitivity and a moderate FDR in a simulation scenario [119].

The GUESS algorithm is a Bayesian variable selection technique and is a search algorithm based on a multichain genetic algorithm. It was developed to explore complex genetic association models and maximize the detection of genetic variation [126]. Similar to the DSA algorithm, the GUESS algorithm showed high sensitivity and a low FDR in simulation research [119]. However, to explain confounding factors, the results must first be fitted to these confounding factors, and the GUESS model then fitted to the residuals.

**Dimension reduction techniques**

Dimensional reduction techniques such as factor analysis or principal component analysis (PCA) have been used to analyze the impact of multiple pollutants on the risk of age-related diseases [127–129]. PCA uses the eigenvalues and eigenvectors of the exposure correlation matrix, makes full use of data variability, and derives orthogonal components from the set of exposure variables. The resulting “feature exposure” can be used for subsequent analysis. However, the relationship PCA determines between exposure and response variables does not take into account the generation of principal components. An improved version of PCA, supervised principal component analysis, can solve this problem [130], as it eliminates the prescreening results of pollutants that are not related to the result and thereby generates an estimated effect with a smaller deviation than the corresponding PCA estimate. The PISCINA-II study (part of the EXPOsOMICS project) aimed to apply the partial least squares (PLS) regression method to investigate the effect of disinfectant product exposure on inflammation [115, 131]. Sparse PLS (sPLS) regression has previously been used in research, such as to identify multiple pollutant exposure profiles related to male reproductive function biomarkers [132]. This method constructs latent variables (linear combinations of predictors) in a supervised manner, that is, uses the results and then regresses the results on the latent variables. The sPLS component not only captures as much predictor variance as possible but also focuses on the variance associated with the results of interest. In a simulation setting, this method showed high sensitivity and a moderate FDR [119].

**Interaction detection models**

In a recent paper, Barrera-Gómez considered scenarios with statistical interactions and systematically compared methods recommended for search interactions [119]. The simulation used contained 237 exposures with real structures, and several statistical regression methods were used. These methods were the two-step full environment association study, the DSA algorithm, the least absolute shrinkage and selection operator (LASSO), and the group-LASSO interaction network (GLINTERNET), which is a three-step method based on a regression tree and an enhanced regression tree. The GLINTERNET and DSA methods performed equally well in terms of the ability to capture interactive items but exhibited an identical trade-off between sensitivity and FDR.

The Bayesian model average (BMA) method is another statistical method used for the construction of a health
risk model that considers multiple pollutants and their interactions. The BMA method is effective for dealing with model uncertainty, as it provides robust estimates of parameters through model averaging. That is, compared with traditional modeling methods that ignore model uncertainties, the BMA method is attractive because it does not choose a single “best” model but instead makes average inferences within a range of possible models [133, 134]. In a simulation study by Sun et al., the BMA method proved useful for the selection of variables with moderately strong exposure–response associations [135].

Mixture approaches
Notably, most of the current chemical exposure–health effect analyses are based on the risk assessment of individual compounds rather than chemical mixtures. However, even exposure to a single pollutant or dose might not cause detrimental health effects, the simultaneous and cumulative exposure to a variety of pollutants may induce the adverse health outcomes. For mixture effect analysis, numerous statistical approaches, such as weighted quantile sum (WQS) regression, Bayesian kernel machine regression (BKMR), and quantile G-computation, have been developed to interpret the overall effect of chemical coexposure on health outcomes and the relative importance of each exposure [136–138]. Recently, Caporale et al. used data from the Swedish Environmental Longitudinal, Mother and child, Asthma and allergy (SELMA) pregnancy cohort and developed a mixture-centric risk assessment strategy integrating epidemiological and experimental data [139]. By applying WQS regression to establish associations between endocrine-disrupting chemical (EDC) mixture exposure and language delay in children, they found that exposure to EDC mixtures in early pregnancy was associated with language delay in offspring. This study highlights the need to consider mixtures in chemical testing and risk assessment processes and provides a comprehensive framework to guide risk assessment strategies.

Overall, the above studies have shown that systems approaches are needed to integrate diverse information in exposomic research (Fig. 1). The systematic evaluation of the relationship between exposures and associated biological responses, such as epigenetic modifications and metabolites, provides valuable insights into key elements of the exposome and delineates how environmental factors and their interactions contribute to the risk of age-related diseases. The statistical strategy in this framework usually contains (I) an EWAS model to explore the associations of exposure–health outcomes; (II) variable selection tools as well as dimension reduction techniques for exposure screening and feature dimension reduction; (III) interaction detection models as well as mixture approaches for assessing the mixture effect, relative contribution, and interactions of chemical coexposure; and (IV) a multiomics network linking exposure to disease phenotype by elucidating their underlying biological mechanisms.

Challenges and opportunities
Multidisciplinary approaches
Each of the subdisciplines of environmental health science can aid in the elucidation of the connection between the exposome and age-related diseases. Information and findings from geriatric populations (via geriatric epidemiology), accurate measurements of chemical species in the environment (via exposure science), exposome–health outcome association analysis (via biostatistics), and data on biological mechanisms and target pathways (via toxicology, molecular biology, and bioinformatics) will all be needed for exposomic research. The findings from these major subdisciplines will need to be integrated via computational methods, together with data from analytical chemistry, genomics, behavioral science, nutritional sciences, and many other fields.

Environmental health science is most powerful when it capitalizes on the transfer of knowledge between its three major subdisciplines: toxicology, geriatric epidemiology, and exposure science. Toxicologists rely on geriatric epidemiological studies and the analysis of exposures to determine which compounds linked to age-related diseases should be studied. In turn, geriatric epidemiologists rely on toxicologists to determine if their observed associations are in line with what is known about the biological and toxicological pathways involved in the aging process or age-related diseases. Toxicologists also rely on exposure scientists to determine what levels of exposure are relevant. If data from all three subdisciplines agree that a particular environmental exposure has an impact on the risk of age-related diseases, this is very strong scientific evidence that such an impact is real. A similar level of convergence among all components of environmental health sciences will be necessary for data from the exposome to have an impact on studies of the risks of age-related diseases.

Genome–exposome epidemiology
The gene–environment (G × E) interaction is a key driver of human health and age-related diseases. The genetic coding system that has evolved to allow individuals to adapt to the environment is supported by a memory system to improve survival and reproduction [140, 141]. Ultimately, genetic epidemiology and exposure epidemiology must fully incorporate this interaction to account for the complex accumulation of G × E interactions over an individual’s lifetime. Given that the exposome is the
accumulation of environmental impacts and related biological responses throughout the entire lifespan [76], predictive models of health outcomes and age-related diseases must comprehensively understand the genetic memory system, as it enables individual genomes to learn from their exposure history and thereby improve their response to subsequent exposures.

Exposure epidemiology is a necessary condition to solve the $G \times E \times I$ (again, with $G$ and $E$ are functions of $I$) integral, which is a function of $E$ and is thus constantly changing throughout an individual's life. However, when the number of interacting elements exceeds $10^{18}$ (20,000 genes × 20,000 splicing variant transcripts × 100,000 or more proteins × 20 or more posttranslational modifications × 10,000 or more metabolites), it is impractical to conduct a systematic investigation of all assumptions [142]. One solution would be to obtain an informative longitudinal analysis of personal genetic and exposure information, as although the initial power would be limited, this would provide a platform to achieve long-term goals and clarify the level of the population whose exposure affects the results [143].

When discussing the analytical complexity inherent to determining an underlying $G \times E$ interaction, it should be noted that a statistical interaction does not represent a biological interaction [144]. However, by using emerging and abundant biological data sources, other potential signals can be identified from the large space of potential interactions [145]. Recently, sophisticated methods have been developed for exploring large-scale $G \times E$ interactions, and these methods could help solve the problem of dealing with many variables [9, 146].

At present, a few studies have examined interactions between sequence variations in the whole genome and environmental exposure to identify novel genetic modulators for certain environmental factors in age-related diseases. For instance, Biernacka et al. performed the first genome-wide gene–environment interaction analysis of pesticide exposure and risk of PD using data on > 700,000 single nucleotide polymorphisms (SNPs) in 364 discordant sibling pairs [147]. They applied a PCA and logistic regression model to detect SNP–pesticide interactions and found that the effect of pesticide exposure on PD risk may be modified by SNPs of the ERCC6L2 gene. Patel et al. used the NHANES and screened 18 SNPs and 5 serum-based environmental factors (results from previous GWASs and EWWs) for interactions in association with T2DM [148]. The results showed that the interaction between rs13266634 (SLC30A8) and trans-β-carotene withstood Bonferroni correction ($P = 0.006$), and the per risk allele effect sizes among subjects with low levels of trans-β-carotene were 40% greater than the marginal effect size (OR: 1.8; 95% CI 1.3–2.6). Nonetheless, understanding the complex interplay between genes and the environment based on whole genome-wide and whole exposome-wide designs and more sophisticated models that incorporate the complexity of exposure mixtures, multiple variants, and multiple types of genomic data (epigenetic, genetic, transcriptomic, etc.) will likely shed new insights on this issue.

**Perspectives and conclusions**

The development of the exposome concept and related technologies will have several impacts on precise public health research and delivery.

1. The first impact will concern vulnerable populations, as the exposome offers an overall view and integrates the source of such populations’ vulnerability with the other exposures that an individual may have and which can aggravate or cause a pathological condition. These vulnerabilities can relate to life stage (e.g., affect children or older adults), sex, genetic factors, dietary origin, pathologies, or particular socioeconomic conditions [149, 150]. For example, older adults are generally considered a susceptible population because of the gradual decline in physiological processes over time. In addition, dosimetric studies showed that there was reduced clearance of PM in all regions of the respiratory tract with increasing age beyond young adulthood [151].

2. The second impact will concern the communication and dissemination of public health messages. Such messages regarding age-related diseases can be contradictory, e.g., “eat more vegetables” but “avoid exposure to pesticides.” Thus, a more integrated approach would give more credibility to these messages.

3. The third impact will concern perspective data reflecting exposures, which can be incorporated into precision medicine to complement genetic information. At a large scale, a systems biology approach can integrate various levels of heterogeneous information. This could include both genetic data and exposomic data, including multiomics data, thereby facilitating a more comprehensive analysis of disease etiologies and specific vulnerabilities than has been provided by previous studies. The subsequent challenge will be to more broadly integrate various risk exposures to obtain a global and more realistic overview of exposure–genetic interactions [152–154].

4. The last impact will involve the bench-to-bedside translation of research. The increasing combination of exposomic data with multiomics data will enable...
researchers to readily identify interim biomarkers of exposure and response that will reveal subtle biological changes. These will facilitate early diagnosis and the identification of potential intervention targets for the prevention and treatment of age-related diseases. In addition, primary prevention to reduce key exposures would reduce the contribution of environmental stressors to the development of age-related diseases and minimize the adverse effects of chemical stressors on individuals as they age.

Conclusions
The explosion in exposome research related to aging and the risk of age-related diseases may provide an answer—from an environmental perspective—to the question “Can we stop ourselves from aging?”, as recently posed by Science magazine.

Abbreviations
GBD: Global Burden of Diseases, Injuries, and Risk Factors Study; SDI: Socio-demographic index; DALYs: Disability-adjusted life years; PD: Parkinson’s disease; AD: Alzheimer’s disease; COPD: Chronic obstructive pulmonary disease; CAD: Coronary artery disease; T2DM: Type 2 diabetes mellitus; GWAS: Genome-wide association study; HR: Hazard ratio; OR: Odds ratio; RR: Relative risk; CI: Confidence interval; PM2.5: 2.5-µM particulate matter; PM10: 10-µM particulate matter; CO: Carbon monoxide; SO2: Sulfur dioxide; NO2: Nitrogen dioxide; Pb: Lead; Al: Aluminum; PFAS: Per- and polyfluoroalkyl substances; dB(A): A-weighted decibels; GIS: Geographic information system; HRMS: High-resolution mass spectrometry; EWAS: Exposome-wide association study; FDR: False discovery rate; DSA: Deletion/substitution/addition; ENET: Elastic net; GUESS: Graphical unit evolutionary stochastic search; PCA: Principal component analysis; PLS: Partial least squares; sPLS: Sparse PLS; LASSO: Least absolute shrinkage and selection operator; GLINTERNET: Group-LASSO interaction network; BMA: Bayesian model average; GxE: Gene–environment interaction.

Supplementary Information
The online version contains supplementary material available at https://doi.org/10.1186/s40246-022-00428-6.

Additional file 1. The search strategies and main results of epidemiologic studies reporting the association between environmental risk factors and age-related diseases.

Acknowledgements
Not applicable.

Author contributions
All authors participated in this work. EMD, YW, and JL were involved in search strategy and drafting. ST and XMS revised and edited the draft and also supervised the project. All authors read and approved the final manuscript.

Funding
This study was financially supported by the National Research Program for Key Issues in Air Pollution Control of China (No. DQGG0401), the National Natural Science Foundation of China (Nos. 82025030, 81941023, and 92043301), and the National Key Research and Development Program of China (No. 2016YFC0206500) to Prof. Xiaoming Shi.

Availability of data and materials
Data sharing is not applicable to this article as no datasets were generated or analyzed during the current study.

Declarations
Ethics approval and consent to participate
Not applicable.

Consent for publication
Not applicable.

Competing interests
None of the authors have any conflict of interest to declare.

Author details
1 China CDC Key Laboratory of Environment and Population Health, National Institute of Environmental Health, Chinese Center for Disease Control and Prevention, Beijing 100021, China. 2 Jiangsu Province Center for Disease Prevention and Control, Nanjing 210009, Jiangsu, China. 3 Center for Global Health, School of Public Health, Nanjing Medical University, Nanjing 211166, Jiangsu, China.

Received: 28 June 2022   Accepted: 29 October 2022
Published online: 08 November 2022

References
1. Campisi J, Kapahi P, Lithgow GJ, Melov S, Newman JC, Verdin E. From discoveries in ageing research to therapeutics for healthy ageing. Nature. 2019;571(7764):89–92.
2. Chang AT, Skirbekk V, Tyrovolas S, Kassebaum NJ, Dieleman JL. Measuring population ageing: an analysis of the Global Burden of Disease Study 2017. Lancet Public Health. 2019;4(3):e159–67.
3. Erikson GA, Bodian DL, Rueda M, Molpiaia B, Scott ER, Scott-Van Zeeland AA, Topol SE, Wineinger NE, Niederhuber JE, Topol EJ, et al. Whole-genome sequencing of a healthy aging cohort. Cell. 2016;165(4):1002–11.
4. Rossiello F, Jurk D, Passos JF, Adda di Fagagna F. Telomere dysfunction in aging and age-related diseases. Nat Cell Biol. 2022;24(2):135–47.
5. de Magalhaes JP, Stevens M, Thornton D. The business of anti-aging science. Trends Biotechnol. 2017;35(11):1062–73.
6. Chakravarti A, Little P. Nature, nurture and human disease. Nature. 2003;421(6921):412–4.
7. Hemminki K, Bermejo JL, Forst A. The balance between heritable and environmental aetiology of human disease. Nat Rev Genet. 2006;7(12):958–65.
8. Lim SS, Vos T, Flaxman AD, Danaei G, Shibuya K, Adair-Rohani H, Amann M, Anderson HR, Andrews KG, Aryee M, et al. A comparative risk assessment of burden of disease and injury attributable to 67 risk factors and risk factor clusters in 21 regions, 1990–2010: a systematic analysis for the Global Burden of Disease Study 2010. Lancet. 2012;380(9859):2223–60.
9. Rappaport SM. Genetic factors are not the major causes of chronic diseases. PLoS ONE. 2016;11(4):1–9.
10. Manolio TA, Collins FS, Cox NJ, Goldstein DB, Hindorff LA, Hunter DJ, McCarthy MI, Ramos EM, Cardon LR, Chakravarti A, et al. Finding the missing heritability of complex diseases. Nature. 2009;461(7265):747–53.
11. Rappaport SM. Implications of the exposome for exposure science. J Expo Sci Environ Epidemiol. 2011;21(1):5–9.
12. Wild CP. Complementing the genome with an “exposome”: the outstanding challenge of environmental exposure measurement in molecular epidemiology. Cancer Epidemiol Biomark Prev. 2005;14(8):1847–50.
13. Niedzwiecki MW, Walker DJ, Vermeulen R, Chadeau-Hyam M, Jones DP, Miller GW. The exposome: molecules to populations. Annu Rev Pharmacol Toxicol. 2019;59:107–27.
14. Wild CP. The exposome: from concept to utility. Int J Epidemiol. 2012;41(1):24–32.
15. Shin S, Burnett RT, Kwong JC, Hystad P, van Donkelaar A, Brook JR, Copes R, Tu K, Goldberg MS, Villeneuve PJ, et al. Effects of ambient air pollution on incident Parkinson's disease in Ontario, 2001 to 2013: a population-based cohort study. Int J Epidemiol. 2018;47(6):2036–48.

16. Ouinn A, Forsberg B, Adolffson AN, Lind N, Modig L, Nordin M, Nordin S, Adolffson R, Nilsson LG. Traffic-related air pollution and dementia incidence in Northern Sweden: a longitudinal study. Environ Health Perspect. 2016;124(3):306–12.

17. Chang KH, Chang MY, Muo CH, Wu TN, Chen CY, Kao CH. Increased risk of dementia in patients exposed to nitrogen dioxide and carbon monoxide: a population-based retrospective cohort study. PLoS ONE. 2014;9(8):e103078.

18. Han C, Oh J, Lim YH, Kim S, Hong YC. Long-term exposure to fine particulate matter and development of chronic obstructive pulmonary disease in the elderly. Environ Int. 2020;143:105895.

19. Li J, Sun S, Tang R, Qiu H, Huang Q, Mason TG, Tian L. Major air pollutants and risk of COPD exacerbations: a systematic review and meta-analysis. Int J Chron Obstruct Pulmon Dis. 2016;11:3079–91.

20. Coogan PF, White LF, Jerrett M, Brook RD, Su JG, Seto E, Burnett R, Petousi N, Leiter J, et al. Analysis of incident Parkinson's disease: results from the CALI cohort study. Environ Health Perspect. 2011;119(8):1273–9.

21. Coogan PF, White LF, Jerrett M, Brook RD, Su JG, Seto E, Burnett R, Petousi N, Leiter J, et al. Analysis of incident Parkinson's disease: results from the CALI cohort study. Environ Health Perspect. 2011;119(8):1273–9.

22. Coogan PF, White LF, Jerrett M, Brook RD, Su JG, Seto E, Burnett R, Petousi N, Leiter J, et al. Analysis of incident Parkinson's disease: results from the CALI cohort study. Environ Health Perspect. 2011;119(8):1273–9.

23. Coogan PF, White LF, Jerrett M, Brook RD, Su JG, Seto E, Burnett R, Petousi N, Leiter J, et al. Analysis of incident Parkinson's disease: results from the CALI cohort study. Environ Health Perspect. 2011;119(8):1273–9.

24. Coogan PF, White LF, Jerrett M, Brook RD, Su JG, Seto E, Burnett R, Petousi N, Leiter J, et al. Analysis of incident Parkinson's disease: results from the CALI cohort study. Environ Health Perspect. 2011;119(8):1273–9.

25. Coogan PF, White LF, Jerrett M, Brook RD, Su JG, Seto E, Burnett R, Petousi N, Leiter J, et al. Analysis of incident Parkinson's disease: results from the CALI cohort study. Environ Health Perspect. 2011;119(8):1273–9.

26. Coogan PF, White LF, Jerrett M, Brook RD, Su JG, Seto E, Burnett R, Petousi N, Leiter J, et al. Analysis of incident Parkinson's disease: results from the CALI cohort study. Environ Health Perspect. 2011;119(8):1273–9.

27. Coogan PF, White LF, Jerrett M, Brook RD, Su JG, Seto E, Burnett R, Petousi N, Leiter J, et al. Analysis of incident Parkinson's disease: results from the CALI cohort study. Environ Health Perspect. 2011;119(8):1273–9.

28. Coogan PF, White LF, Jerrett M, Brook RD, Su JG, Seto E, Burnett R, Petousi N, Leiter J, et al. Analysis of incident Parkinson's disease: results from the CALI cohort study. Environ Health Perspect. 2011;119(8):1273–9.

29. Coogan PF, White LF, Jerrett M, Brook RD, Su JG, Seto E, Burnett R, Petousi N, Leiter J, et al. Analysis of incident Parkinson's disease: results from the CALI cohort study. Environ Health Perspect. 2011;119(8):1273–9.

30. Coogan PF, White LF, Jerrett M, Brook RD, Su JG, Seto E, Burnett R, Petousi N, Leiter J, et al. Analysis of incident Parkinson's disease: results from the CALI cohort study. Environ Health Perspect. 2011;119(8):1273–9.

31. Coogan PF, White LF, Jerrett M, Brook RD, Su JG, Seto E, Burnett R, Petousi N, Leiter J, et al. Analysis of incident Parkinson's disease: results from the CALI cohort study. Environ Health Perspect. 2011;119(8):1273–9.

32. Coogan PF, White LF, Jerrett M, Brook RD, Su JG, Seto E, Burnett R, Petousi N, Leiter J, et al. Analysis of incident Parkinson's disease: results from the CALI cohort study. Environ Health Perspect. 2011;119(8):1273–9.

33. Coogan PF, White LF, Jerrett M, Brook RD, Su JG, Seto E, Burnett R, Petousi N, Leiter J, et al. Analysis of incident Parkinson's disease: results from the CALI cohort study. Environ Health Perspect. 2011;119(8):1273–9.

34. Coogan PF, White LF, Jerrett M, Brook RD, Su JG, Seto E, Burnett R, Petousi N, Leiter J, et al. Analysis of incident Parkinson's disease: results from the CALI cohort study. Environ Health Perspect. 2011;119(8):1273–9.

35. Coogan PF, White LF, Jerrett M, Brook RD, Su JG, Seto E, Burnett R, Petousi N, Leiter J, et al. Analysis of incident Parkinson's disease: results from the CALI cohort study. Environ Health Perspect. 2011;119(8):1273–9.
53. Franceschi C, Garagnani P, Parini P, Giuliani C, Santoro A. Inflammaging: a new immune-metabolic viewpoint for age-related diseases. Nat Rev Endocrinol. 2018;14(10):576–90.

54. Savolainen-Peltokangas H, Rahkola-Sosiala P, Hott F, Vattulainen P, Pissler M, Ylikahla O, Mikkola TS. Use of postmenopausal hormone therapy and risk of Alzheimer's disease in Finland: nationwide case-control study. BMJ. 2019;364:k3665.

55. Wu M, Li M, Yuan J, Liang S, Chen Z, Ye M, Ryan PM, Clark C, Tan SC, Rahmani J, et al. Postmenopausal hormone therapy and Alzheimer's disease, dementia, and Parkinson's disease: a systematic review and time-response meta-analysis. Pharmacol Res. 2020;155:104693.

56. Wu YT, Prina AM, Jones AP, Barnes LE, Matthews FE, Brayne C, Medical Research Council Cognitive F, Ageing S, Community environment, cognitive impairment and dementia in later life: results from the cognitive function and ageing study. Age Ageing. 2015;44(6):1005–11.

57. de Keijzer C, Tonne C, Basagana X, Valentin A, Singh-Manoux A, Alonso A, Wijmenga N, Ovsiannikov A, Elliott P, et al. Social integration, social support, and all-cause, cardiovascular disease and cause-specific mortality: a prospective cohort study. Int J Environ Res Public Health. 2019;16(9):1498.

58. Sarkar C, Zhang B, Ni M, Kumari S, Bauermeister S, Gallacher J, Webster C. Environmental correlates of chronic obstructive pulmonary disease in 96,779 participants from the UK Biobank: a cross-sectional, observational study. Lancet Planet Health. 2019;3(11):e478–90.

59. Tan J, Wang Y. Social integration, social support, and all-cause, cardiovascular disease and cause-specific mortality: a prospective cohort study. Int J Environ Res Public Health. 2019;16(9):1498.

60. Penninkinkamp R, Casey AN, Singh MF, Brodaty H. The association between social engagement, loneliness, and risk of dementia: a systematic review and meta-analysis. J Alzheimers Dis. 2018;66(4):1619–33.

61. Jia L, Du Y, Chu L, Zhang Z, Lu F, Lyu D, Li Y, Li Y, Zhu M, Jiao H, et al. Prevalence, risk factors, and management of dementia and mild cognitive impairment in adults aged 60 years or older in China: a cross-sectional study. Lancet Public Health. 2020;5(2):e661–71.

62. Petersen JD, Wehberg S, Packness A, Svensson NH, Hyldig N, Rungsgaard S, Andersen MK, Ryg J, Mercer SW, Sondergaard J, et al. Associations between neighborhood physical and social environments and incident type 2 diabetes mellitus: the multi-ethnic study of atherosclerosis (MESA). JAMA Intern Med. 2015;175(18):2693–701.

63. Kennedy BK, Berger SL, Brunet A, Campisi J, Cuervo AM, Epel ES, Franceschi C, Lithgow GJ, Morimoto RI, Pessin JE, et al. Geroscience: linking marks of aging. Cell. 2013;153(6):1194–217.

64. Lopez-On tin C, Blasco MA, Partridge L, Serrano M, Kroemer G. The hallmarks of aging. Cell. 2014;159(4):709–13.

65. Martin EM, Fry RC. Environmental influences on the epigenome: exposure-associated DNA methylation in human populations. Annu Rev Public Health. 2018;39:309–33.

66. Huat TJ, Camats-Perma J, Newcombe EA, Valmas N, Kitazawa M, Medeiros R. Metal toxicity links to Alzheimer's disease and neuroinflammation. J Mol Biol. 2019;431(19):1843–68.

67. Martin EM, Fry RC. Environmental influences on the epigenome: exposure-associated DNA methylation in human populations. Annu Rev Public Health. 2018;39:309–33.

68. Roy AL, Sierra F, Howcroft K, Singer DS, Sharpless N, Hodes RJ, Wilder EL, Anderson JM. A blueprint for characterizing senescence. Cell. 2020;183(5):1143–6.

69. Kubi JA, Chen ACH, Fong SW, Lai KP, Wong CKC, Lee KF, Soundararajan A, Prabu P, Mohan V, Gibert Y, Balasubramanyam SG, Cavalli G, Heard E. Advances in epigenetics link genetics to the environment. Cell. 2019;183(5):1143–6.

70. Lehallier B, Gate D, Schaum N, Nanasi T, Lee SE, Youssef H, Moran Losada P, Berndik D, Keller A, Vergheese J, et al. Undulating changes in human plasma proteome profiles across the lifespan. Nat Med. 2020;25(12):1843–50.

71. Paul KC, Sinhaheiser JS, Rhodes SL, Cockburn M, Bronstein J, Ritz B. Organophosphate pesticide exposures, nitric oxide synthase gene variants, and pesticide interactions in a case-control study of Parkinson's disease, California (USA). Environ Health Perspect. 2016;124(5):570–7.

72. Christine PJ, Auckincloss AH, Berntini AG, Cariello MR, Schnetz BN, Moore K, Kard JE, Horwich TB, Watson KE, Rox AE. Longitudinal associations between neighborhood physical and social environments and incident type 2 diabetes mellitus: the multi-ethnic study of atherosclerosis (MESA). JAMA Intern Med. 2015;175(18):2693–701.

73. Haddad N, Andriaou XD, Makris KC. A scoping review on the characteristics of human exposome studies. Curr Pollut Rep. 2019;5(4):378–93.

74. Rappaport SM, Smith MT. Epidemiology, environment and disease risks. Science. 2010;330(6003):460–6.

75. Dennis KS, Auerbach SS, Balbask DM, Cui Y, Fallin MD, Smith MT, Spira A, Summer S, Miller GW. The importance of the biological impact of exposure to the concept of the exposome. Environ Health Perspect. 2016;124(10):1504–10.

76. Miller GW, Jones DP. The nature of nurture: refining the definition of the exposome. Toxicol Sci. 2014;137(1):1–2.

77. Wagner KH, Cameron-Smith D, Wessner B, Franke B. Biomarkers of aging: from function to molecular biology. Nutrients. 2016;8(6):338.

78. Rivero-Segura NA, Bello-Chavolla OY, Barrera-Vazquez OS, Gutierrez-Robledo LM, Gomez-Vejaran JC. Promising biomarkers of human aging: In search of a multi-omics panel to understand the aging process from a multidimensional perspective. Ageing Res Rev. 2020;64:101164.

79. Fang M, Hu L, Chen D, Guo Y, Liu J, Lan C, Gong J, Wang B. Exposure to pesticides and all-cause mortality: a systematic review and meta-analysis. J Alzheimers Dis. 2018;66(4):1619–33.

80. Petersen JD, Wehberg S, Packness A, Svensson NH, Hyldig N, Rungsgaard S, Andersen MK, Ryg J, Mercer SW, Sondergaard J, et al. Associations between neighborhood physical and social environments and incident type 2 diabetes mellitus: the multi-ethnic study of atherosclerosis (MESA). JAMA Intern Med. 2015;175(18):2693–701.

81. Turner MC, Nieuwenhuijsen M, Anderson K, Balshaw D, Cui Y, Dunton G, Hoppin JA, Kouraklis P, Jrettet M. Assessing the exposome with external measures: commentary on the state of the science and research recommendations. Annu Rev Public Health. 2017;38:215–39.

82. Geddes JA, Martin RV, Boys BL, van Donkelaar A. Long-term trends in TSP and ambient OA concentrations. Environ Health Perspect. 2016;124(3):281–9.

83. Nieuwenhuijsen MJ, Kruize H, Gidlow C, Andrusiayte S, Anto JM, Basagana X, Cirach M, Dadvand P, Danileviciute A, Donaire-Gonzalez D, et al. Positive health effects of the natural outdoor environment in people with diabetes. BMJ Open. 2014;4(4):e004951.
van Donkelaar A, Martin RV, Brauer M, Kahn R, Levy R, Verduzco C, Villeneuve PJ. Global estimates of ambient fine particulate matter concentrations from satellite-based aerosol optical depth: development and application. Environ Health Perspect. 2010;118(6):847–55.

Brook RD, Cakmak S, Turner MC, Brook JR, Crouse DL, Peters PA, van Donkelaar A, Villeneuve PJ, Brion O, Jerrett M, et al. Long-term fine particulate matter exposure and mortality from diabetes in Canada. Diabetes Care. 2013;36(10):3313–20.

course DL, Peters PA, van Donkelaar A, Goldberg MS, Villeneuve PJ, Brion O, Khan S, Atani DO, Jerrett M, Pope CA, et al. Risk of nonaccidental and cardiovascular mortality in relation to long-term exposure to low concentrations of fine particulate matter: a Canadian national-level cohort study. Environ Health Perspect. 2012;120(5):708–14.

Lee M, Klooq I, Chudnovsky A, Wang Y, Melly S, Coull B, Koutrapis P, Schwartz J. Spatiotemporal prediction of fine particulate matter using high-resolution satellite images in the Southeastern US 2003–2011 (Review). J Exp Sci Environ Epidemiol. 2016;26(4):377–84.

course DL, Peters PA, van Donkelaar A, Goldberg MS, Villeneuve PJ, Brion O, Khan S, Atani DO, Jerrett M, Pope CA, et al. Risk of nonaccidental and cardiovascular mortality in relation to long-term exposure to low concentrations of fine particulate matter: a Canadian national-level cohort study. Environ Health Perspect. 2012;120(5):708–14.

Moeller R, De Boever P Perspectives for environment and health research in Horizon 2020: dark ages or golden era? Int J Hyg Environ Health. 2014;217(4):889–91.

Bailey E, Fuhrmann C, Runkle J, Stevens S, Brown M, Sugg M. Wearable sensors for personal temperature exposure assessments: a comparative study. Environ Res. 2020;180:108858.

Constantinou A, Okonomou S, Konstantinou C, Makris KC. A randomized cross-over trial investigating differences in 24-hour personal air and skin temperatures using wearable sensors between two climatologically contrasting settings. Sci Rep. 2021;11(1):22020.

Gao P, Shen X, Zhang X, Jiang C, Zhang S, Zhou X, Schussler-Fiorenza Rose SM, Snyder M. Precision environmental health monitoring by longitudinal exposome and multi-omics profiling. Genome Res. 2022;32(6):1199–214.

Durrin J, Kozdinski J, Runge R, Witt D, Hicks JI, Schussler-Fiorenza Rose SM, Xi L, Bahmani A, Delp SL, Hastie T, et al. Wearable sensors enable personalized predictions of clinical laboratory measurements. Nat Med. 2021;27(6):1105–12.

Koelmel JP, Lin EZ, Guo P, Zhou J, He J, Chen A, Gao Y, Deng F, Dong H, Liu Y, Cha Y, Fang J, Beecher C, Shi X, Tang S, Godd Pollit KJ. Exploring the external exposome using wearable passive samplers—the China BAPE study. Environmental pollution (Barking, Essex: 1987). 2020;270:116228.

Tang S, Li T, Fang J, Chen R, Cha Y, Wang Y, Zhu M, Zhang Y, Chen Y, Du Y, et al. The exposome in practice: an exploratory panel study of biomarkers of air pollutant exposure in Chinese people aged 60–69 years (China BAPE Study). Environ Res. 2021;157:106866.

Jiang C, Wang X, Li X, Inlora J, Wang T, Li Q, Snyder M. Dynamic human environmental exposome revealed by longitudinal personal monitoring. Cell. 2018;175(1):277-291 e231.

Jiang C, Zhang X, Gao P, Chen Q, Snyder M. Decoding personal biotic and abiotic airborne exposome. Nat Protoc. 2021;16(2):1129–51.

Schymanski EL, Singer HP, Longpre P, Loos M, Ruff M, Stravs MA, Ripollès Vidal C, Hollender J. Strategies to characterize polar organic contamination in wastewater: exploring the capability of high resolution mass spectrometry. Environ Sci Technol. 2014;48(3):1811–8.

Sevin DC, Kuehne A, Zamboni N, Sauer U. Biological insights through metabolomics. Environ Sci Technol. 2014;48(3):1811–8.

Arif AA, Shah SM. Association between personal exposure to volatile organic compounds and asthma among US adult population. Int Arch Occup Environ Health. 2007;80(8):711–9.

Roberts S, Martin MA. Using supervised principal components analysis to assess multiple pollutant effects. Environ Health Perspect. 2006;114(12):1877–82.
131. Jain P, Vineis P, Liquet B, Vlaanderen J, Bodinier B, van Veldhoven K, Kogevinas M, Athersuch TJ, Font-Ribera L, Villanueva CM, et al. A multivariate approach to investigate the combined biological effects of multiple exposures. J Epidemiol Community Health. 2018;72(7):564–71.

132. Lenters V, Portengen L, Smit LA, Jonsson BA, Givwercman A, Rylander L, Lindh CH, Spano M, Pedersen HS, Ludwicki JK, et al. Phthalates, perfluoroalkyl acids, metals and organochlorines and reproductive function: a multipollutant assessment in Greenlandic, Polish and Ukrainian men. Occup Environ Med. 2015;72(6):385–93.

133. Thomas DC, Jerrett M, Kuenzli N, Louis TA, Dominici F, Zeger S, Schwartz J, Burnett RT, Krewski D, Bates D. Bayesian model averaging in time-series studies of air pollution and mortality. J Toxicol Environ Health A. 2007;70(3–4):311–5.

134. Thomas DC, Witte JS, Greenland S. Dissecting effects of complex mixtures: who’s afraid of informative priors? Epidemiology. 2007;18(2):186–90.

135. Sun Z, Tao Y, Li S, Ferguson KK, Meeker JD, Park SK, Batterman SA, Mukherjee B. Statistical strategies for constructing health risk models with multiple pollutants and their interactions: possible choices and comparisons. Environ Health. 2013;12(1):85.

136. Carrico C, Gennings C, Wheeler DC, Factor-Litvak P. Characterization of weighted quantile sum regression for highly correlated data in a risk analysis setting. J Agric Biol Environ Stat. 2015;20(1):100–20.

137. Bobb JF, Valen L, Claus Henrik B, Christiani DC, Wright RO, Mazumdar M, Godleski JJ, Coull BA. Bayesian kernel machine regression for estimating the health effects of multi-pollutant mixtures. Biostatistics. 2015;16(3):493–508.

138. Keil AP, Buckley JP, O’Brien KM, Ferguson KK, Zhao S, White AJ. A quantile-based g-computation approach to addressing the effects of exposure mixtures. Environ Health Perspect. 2020;128(4):47004.

139. Caporale N, Leemans M, Birgersson L, Germain PL, Cheroni C, Borbely G, Engdahl E, Lindh C, Bressan RB, Cavallo F, et al. From cohorts to molecules: adverse impacts of endocrine disrupting mixtures. Science. 2022;375(6582):eabe8244.

140. Jones DP, Sies H. The redox code. Antioxid Redox Signal. 2015;23(9):734–46.

141. Go YM, Jones DP. Exposure memory and lung regeneration. Ann Am Thorac Soc. 2016;13(Suppl 2):S452–61.

142. Go YM, Fernandes J, Hu X, Uppal K, Jones DP. Mitochondrial network responses in oxidative physiology and disease. Free Radic Biol Med. 2018;116:31–40.

143. Jones DP. Sequencing the exposome: a call to action. Toxicol Rep. 2016;3:29–45.

144. Thomas D. Gene–environment-wide association studies: emerging approaches. Nat Rev Genet. 2010;11(4):259–72.

145. Patel CJ. Analytical complexity in detection of gene variant-by-environment exposure interactions in high-throughput genomic and exposomic research. Curr Environ Health Rep. 2016;3(1):64–72.

146. Westerman KE, Pham DT, Hong L, Chen Y, Sevilla-Gonzalez M, Sung YJ, Sun YY, Morrison AC, Chen H, Manning AK. GEM: scalable and flexible gene–environment interaction analysis in millions of samples. Bioinformatics. 2021;37(20):3514–20.

147. Biermacka JM, Chung SJ, Armamas SM, Anderson KS, Lill CM, Bertram L, Ahlskog JE, Brighina L, Frigenio R, Maraganore DM. Genome-wide gene-environment interaction analysis of pesticide exposure and risk of Parkinson’s disease. Parkinsonism Relat Disord. 2016;32:25–30.

148. Patel CJ, Chen R, Kodama K, Ioannidis JP, Butter AJ. Systematic identification of interaction effects between genome- and environment-wide associations in type 2 diabetes mellitus. Hum Genet. 2013;132(3):495–508.

149. Barouki R, Guclukman PO, Grandjean P, Hansson M, Heindel JJ. Developmental origins of non-communicable disease: implications for research and public health. Environ Health. 2012;11:42.

150. Balbus JM, Barouki R, Birmbaum LS, Etzel RA, Guclukman PO, Sr, Grandjean P, Hancock C, Hansson MA, Heindel JJ, Hoffman K, et al. Early-life prevention of non-communicable diseases. Lancet. 2013;381(9860):3–4.

151. Sacks JD, Stanek IJ, Luben TJ, Johans DO, Buckley BJ, Brown JS, Ross M. Particular matter-induced health effects: who is susceptible? Environ Health Perspect. 2011;119(4):446–54.

152. Barouki R, Audouze K, Coumoul X, Demenais F, Gauguer D. Integration of the human exposome with the human genome to advance medicine. Biochimie. 2018;152:155–8.

153. Wishart DS. Emerging applications of metabolomics in drug discovery and precision medicine. Nat Rev Drug Discov. 2016;15(7):473–84.

154. Khoury MJ, Galea S. Will precision medicine improve population health? JAMA. 2016;316(13):1357–8.

Publisher’s Note
Springer Nature remains neutral with regard to jurisdictional claims in published maps and institutional affiliations.