Summertime partitioning and budget of NO\textsubscript{y} compounds in the troposphere over Alaska and Canada: ABLE 3B

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As part of NASA's Arctic Boundary Layer Expedition 3A and 3B field measurement programs, measurements of NO\textsubscript{x}, HNO\textsubscript{3}, PAN, PPN, and NO\textsubscript{y} were made in the middle to lower troposphere over Alaska and Canada during the summers of 1988 and 1990. These measurements are used to assess the degree of closure within the reactive odd nitrogen (N\textsubscript{X}O\textsubscript{Y}) budget through the comparison of the values of NO measured with a catalytic convertor to the sum of individually measured NO\textsubscript{y}(i) compounds (i.e., ΣNO\textsubscript{y}(i) = NO\textsubscript{x} + HNO\textsubscript{3} + PAN + PPN). Significant differences were observed between the various study regions. In the lower 6 km of the troposphere over Alaska and the Hudson Bay lowlands of Canada a significant fraction of the NO\textsubscript{y} budget (30 to 60%) could not be accounted for by the measured ΣNO\textsubscript{y}(i). This deficit in the NO\textsubscript{y} budget is about 100 to 200 parts per trillion by volume (pptv) in the lower troposphere (0.15 to 3 km) and about 200 to 400 pptv in the middle free troposphere (3 to 6.2 km). Conversely, the NO\textsubscript{y} budget in the northern Labrador and Quebec regions of Canada is almost totally accounted for within the combined measurement uncertainties of NO\textsubscript{x} and the various NO\textsubscript{y}(i) compounds. A substantial portion of the NO\textsubscript{y} budget's "missing compounds" appears to be coupled to the photochemical and/or dynamical parameters influencing the tropospheric oxidative potential over these regions. A combination of factors are suggested as the causes for the variability observed in the NO\textsubscript{y} budget. In addition, the apparent stability of compounds represented by the NO\textsubscript{y} budget deficit in the lower-altitude range questions the ability of these compounds to participate as reversible reservoirs for "active" odd nitrogen and suggest that some portion of the NO\textsubscript{y} budget may consist of relatively unreactive nitrogen-containing compounds.

1. INTRODUCTION

In many regions of the remote troposphere the availability of reactive odd nitrogen in the form of nitric oxide (NO) is a critical factor controlling the photochemical production of ozone (O\textsubscript{3}). This control of O\textsubscript{3} production occurs through NO reaction with hydroperoxy and organoperoxy radicals (HO\textsubscript{2} and RO\textsubscript{2}). The photochemical destruction rate for O\textsubscript{3} is believed to be controlled by reactions of O\textsubscript{3} with the hydroperoxy radical (HO\textsubscript{2}) and by direct photolysis of O\textsubscript{3} leading to hydroxyl radical (OH) production. The absence of NO would result in a net photochemical loss of O\textsubscript{3} throughout the troposphere. Over remote tropical and midlatitude marine environments, observed NO mixing ratios have been small enough (i.e., NO mixing ratios of a few parts per trillion by volume) to yield predictions of net ozone loss throughout most of the middle to lower troposphere in these locations [e.g., Liu et al., 1983; Chameides et al., 1987; Carroll et al.; 1990]. The NO mixing ratios observed in the summertime troposphere (0.15 - 6.2 km) over Alaska were, on average, slightly larger than those observed over the remote lower-latitude marine environments [Sandholm et al., 1992]. The summertime NO mixing ratios over Alaska were still small enough to result in predictions of a net photochemical loss of O\textsubscript{3} in the lower 6-km column. In these cases the O\textsubscript{3} photochemical lifetime was predicted to be nearly equal to the lifetime based on surface deposition. These results also indicated that the O\textsubscript{3} photochemical lifetime was nearly 2.5 times longer than that predicted in the absence of NO [Jacob et al., 1992]. The middle to lower tropospheric summertime measurements of NO made over the Arctic and subarctic regions of both Alaska and Canada fall within the range of mixing ratios (i.e., NO < 200 pptv [Sandholm et al., 1992; Talbot et al., this issue]) that have been characterized as NO-limited environments with regard to production of the oxidants OH and O\textsubscript{3} [cf. Logan et al., 1981; Lin et al., 1988; Hough and Johnson, 1991; Chameides et al., 1992].

Control of NO mixing ratios on the few minutes timescale is exerted via the fast photochemical cycle involving NO reactions with O\textsubscript{3} and peroxyradicals that lead to the formation of nitrogen dioxide (NO\textsubscript{2}) and NO's subsequent reformation via the photolysis of NO\textsubscript{2}. Permanent loss of tropospheric NO is believed to be dominated by the loss of NO\textsubscript{2} through the formation of nitric acid (HNO\textsubscript{3}) via reactions such as OH + NO\textsubscript{2} (+ M) → HNO\textsubscript{3}. This latter reaction is sufficiently fast in the

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summit time lower troposphere (of the order of hours) to significantly deplete both NO and \( \text{NO}_2 \) (\( \text{NO} + \text{NO}_2 = \text{NO}_3 \)) from air parcels within a few days travel time of their \( \text{NO}_x \) sources \[e.g., \text{Logan}, 1983\].

In contrast to the loss of NO\(_x\) represented above by the HNO\(_3\) sink, NO\(_x\) can be converted to a number of stable reservoir compounds such as pernitric acid (H\(\text{NO}_3\)), peroxyacetyl nitrate (PAN), and various other alkyl nitrates (RONO\(_2\)) and NO\(_x\)-containing hydrocarbons. These reservoir compounds can act to transport the NO\(_x\) pool over large distances \[e.g., \text{Curneen}, 1979; \text{Logan et al.}, 1981; \text{Singh and Hanst}, 1981; \text{Singh}, 1987\], and collectively form the dominant component of the "active" odd nitrogen family in some regions of the remote troposphere. Observations made in the middle to lower summertime troposphere over Alaska have indicated that under certain conditions the thermal decomposition of PAN alone could account for the NO\(_x\) abundance in the lower 6-km tropospheric column \[\text{Singh et al.}, 1992b\], whereas in other cases the middle tropospheric (4 - 6 km) abundance of NO\(_x\) may have been controlled by the degradation of other NO\(_x\) reservoir compounds \[\text{Jacob et al.}, 1992\].

The active odd nitrogen compounds discussed above represent a subset of the larger pool of total atmospheric odd nitrogen \( (\text{N}_2\text{O}_5) \), which includes more stable compounds such as HNO\(_3\) and particulate nitrate (p-NO\(_3\)) that are not currently believed to be readily converted back into NO\(_x\) under typical middle to lower tropospheric conditions. Measurement of total odd nitrogen \( (\text{N}_2\text{O}_5) \) is possible using the analytical techniques based on the catalytic reduction of N\(_2\text{O}_5\) compounds to NO \[\text{Bollinger et al.}, 1983; \text{and Fahey et al.}, 1986\]. Several studies have shown that the "total" odd nitrogen compounds measured by this technique \( (\text{N}_2\text{O}_5) \) is often greater than the sum of the subset of individually measurable, NO\(_x\)(i), odd nitrogen compounds \( (e.g., \text{NO}_x > \Sigma \text{NO}_x\text{(i)}) \) \[\text{Fahey et al.}, 1986; \text{Ridley}, 1991; \text{Singh et al.}, 1992; \text{Atlas et al.}, 1992\].

In these studies the subset of individually measured NO\(_x\)(i) compounds was initially believed to represent the bulk of the anticipated N\(_2\text{O}_5\) family members with the list of measured NO\(_x\)(i) compounds increasing in progressive studies \[e.g., \text{Atlas et al.}, 1992; \text{Buhr et al.}, 1990\]. The unidentified N\(_2\text{O}_5\) compounds that are implied from the deficit found in NO\(_x\) - 2NO\(_x\)(i) may serve as reservoirs for the transport of active odd nitrogen into the remote troposphere \[\text{Fahey et al.}, 1986; \text{Singh}, 1987; \text{Roberts}, 1990; \text{Buhr et al.}, 1990; \text{Jacob et al.}, 1992\]. Further characterization of the distribution and magnitude of the known and implied unknown components of the active odd nitrogen reservoir is important to the understanding of the factors controlling the abundance of NO\(_x\) in remote regions.

The characterization of high northern latitude middle to lower tropospheric trace gas distributions was one goal of the NASA Global Tropospheric Experiment (GTE) Arctic Boundary Layer Expedition (ABLE) phase 3 programs. The airborne sampling portions of these programs surveyed the trace gas distributions over Alaska in the summer of 1988 (ABLE 3A) \[e.g., \text{Harriss et al.}, 1992\] and the Hudson Bay lowlands and northeastern boreal forest of Canada in the summer of 1990 (ABLE 3B) \[e.g., \text{Harriss et al.}, this issue\]. A wide range of chemical and physical variables were measured during these programs, including NO, NO\(_2\), PAN, PPN, HNO\(_3\), p-NO\(_3\), O\(_3\), CO, CH\(_4\), nonmethane hydrocarbons (NMHCs), C\(_2\)Cl\(_2\), and H\(_2\)O.

This suite of measurements provides a unique opportunity to study the summertime partitioning of active and reservoir odd nitrogen compounds in the middle to lower tropospheric air masses overlying remote regions of the high-latitude continent of North America. In this paper we examine the partitioning of individual odd nitrogen compounds within the measured N\(_2\text{O}_5\) budget and address issues related to the budget's degree of closure and its relationship with other physical, chemical, and measurement-related variables. Details of the distribution and relationship of odd nitrogen compounds and other trace gases are described in a series of papers presented in the Journal of Geophysical Research special issue for ABLE 3A and in this issue. The entire NASA/GTE ABLE 3 data archive is available from the NASA/GTE project office (Langley, Virginia).

2. MEASUREMENT INSTRUMENTATION

Both the ABLE 3A (Alaskan) and the ABLE 3B (Canadian) airborne field programs were conducted on the NASA Wallops Flight Facility L-188 Electra aircraft. Although some modification of each N\(_2\text{O}_5\) instrument occurred between the 1988 ABLE 3A and the 1990 ABLE 3B programs, the fundamental measurement methodology underlying each instrument was identical for the two programs. Details of the NO\(_x\)(i) and NO\(_x\) measurement techniques are discussed in sections 2.1 through 2.4.

The techniques used for concurrent measurements of O\(_3\), CO, selected nonmethane hydrocarbons, and meteorological variables have been described elsewhere \[\text{Gregory et al.}, 1983; \text{Sachse et al.}, 1987; \text{Blake et al.}, 1992; \text{Harriss et al.}, this issue\].

2.1. NO\(_x\) and NO\(_x\), Measurements

The spectroscopically selective two-photon/laser-induced fluorescence (TP/LIF) NO\(_x\) technique was applied to the simultaneous measurement of NO, NO\(_2\), and NO\(_x\). Details of this instrument have been previously reported \[\text{Bradshaw et al.}, 1985; \text{Sandholm et al.}, 1990; \text{Sandholm et al.}, 1992\]. The 226-nm and 1.1-μm laser beams used in the two-photon fluorescence excitation process were passed through three separate ambient sampling cells. One cell was designated for detecting ambient NO. A second cell was designated for detecting NO produced from a photolytic conversion of ambient NO\(_2\). The third cell was designated for detecting NO produced from the 300°C gold-catalyzed conversion of ambient NO, using 0.3% CO as a reducing agent. The NO, NO\(_2\), and NO\(_x\) channels
used separate signal acquisition electronics, flow measurement and control systems, and signal normalization/internal reference standards.

Ambient air was continuously drawn in through a common porcelain-glass coated inlet (2.5-cm ID) at a nominal flow rate of 200 liters per minute (lpm). The airstream was sampled perpendicular to the aircraft motion. The sample residence time through the inlet manifold was always less than half a second. Total sample residence time through the NO ambient sampling portion of the instrument was less than 1 s. Sample residence time in the NO$_2$ portion of the instrument varied between 1 and 5 s, depending on the photolytic converter system used. NO$_2$ samples were drawn from the center of the ambient sampling manifold at a location near the NO$_2$ converter assembly with a residence time < 0.2 s from the manifold to the NO$_2$ converter. The sampling manifold and flow line fittings were tested in flight for leaks using a 10 parts per million by volume (ppm) NO standard as a leak tracer.

During ABLE 3A an excimer laser was used as the photolytic converter system's photolysis source [Sandholm et al., 1990; Sandholm et al., 1992]. To reduce the instrument's size, weight, and power consumption, a high-pressure xenon arc-lamp-based photolytic converter was used during the ABLE 3B program. This latter converter utilized a 1-kW Cermax short-arc lamp that was spectrally filtered into a photolysis passband of 350 nm < $\lambda$ < 420 nm. This converter was designed to operate at 30% to 60% photolytic efficiency with sample residence times ranging from 2 to 4.5 s, respectively. The low wavelength cutoff (10% peak spectral fluence) of 350 nm was chosen to minimize possible interferences from concomitant R-NO$_x$ compounds. Deep UV ($\lambda$ < 330 nm) and Visible/IR ($\lambda$ > 480 nm) emissions from the lamp were attenuated by > 10$^3$-fold relative to the intensity at the center of the passband. The photolysis beam was also spatially filtered to avoid illumination of the photolytic sample cell's walls. In both the excimer laser and the arc-lamp-based converters, the photolytic sample cells were thermally controlled using high flow rate of ambient air that was passed through an outer jacket of each cell.

The NO$_x$ catalytic converter system used with the TP/LIF sensor was adapted from the NO$_x$ converter system design developed by NOAA [Bollinger et al., 1983; Fahey et al., 1985; Murphy and Fahey, 1987]. In the ABLE 3A program a short length (~ 0.3 m) of PFA Teflon tubing was used to connect the NO$_x$ converter to the main porcelain-glass-coated sampling manifold. This tubing was replaced by a similar length of heated (~40°C) gold-coated nickel tubing in the ABLE 3B program. This change was made to minimize the possible effects of sample line memory during the more rapid descents/ascents planned for the ABLE 3B field program. During the 1989 NASA/GTE Chemical Instrumentation Test and Evaluation (CITE) 3 field program a similar PFA Teflon tubing was used to couple the NO$_x$ converter to the inlet/manifold. In this latter case, a positive artifact was believed to have been induced by sample line memory effects following the rapid descent (500 m/min) through haze layers that occurred near the trade wind inversion over the equatorial western Atlantic Ocean near Natal, Brazil.

Inflight and preflight/postflight calibrations were performed by standard additions to either ambient air (inflight) or bottled air (preflight/postflight). Two-stage serial dilution systems were used to dilute the parts per minute by volume mixing ratios of standards contained by high-pressure aluminum cylinders. The diluted standards were injected directly into the ambient sampling inlet and subsequently diluted to final concentration (0.1 to 2 parts per billion by volume (ppbv)). All flow measurements were made with linear mass flow controllers or meters. These devices were intercompared with positive volume displacement flow measurement instrumentation prior to and after both the ABLE 3A and ABLE 3B field programs.

Primary and secondary gas calibration standards were used for both NO and NO$_2$. The 50 ppmv (in nitrogen) NO and NO$_2$ primary standards used for intercalibration have remained stable since their acquisition in 1981 (i.e., < 10% deviation from initial conditions from 1981 to 1991). Standards were intercompared with National Institute of Standards and Technology (NIST) standards as part of the NASA/GTE field measurement program protocol in 1983, 1984, 1986, 1990, and 1991. High-pressure nitroethane (C$_2$H$_5$NO$_2$) in helium standards were developed to provide a more rigorous test of the NO$_2$ conversion efficiency. These C$_2$H$_5$NO$_2$ standards were made from reagent grade C$_2$H$_5$NO$_2$ that was further purified by several low-temperature vacuum distillations in which only the middle third of the distillate was retained. Aluminum cylinders were vacuum baked (< 10$^3$ mbar, 50°C) and flushed several times, then filled with a known pressure of the purified C$_2$H$_5$NO$_2$ and diluted with research purity helium to a final cylinder pressure of about 1600 psig. Final C$_2$H$_5$NO$_2$ mixing ratios determined from the cylinder contents by UV absorption agreed to within 10% of the value calculated from partial pressure dilution. The C$_2$H$_5$NO$_2$ standards have proven to be stable at mixing ratios in the range of 10 ppmv. Inflight calibrations were made using NO and NO$_2$ during ABLE 3A and using NO, NO$_2$, and C$_2$H$_5$NO$_2$ during ABLE 3B.

The preflight/postflight mixing ratios measured in bottled air to which no standard was added were not subtracted as "blank" values for the NO$_x$ or NO$_2$ measurements reported here. We believe these "blank" values, which ranged from approximately 40 to 120 pptv, predominantly reflected outgassing, under slow flow conditions (≤ 40 slpm), of the long length of extra tubing (~ 15 m) used to reach the aircraft inlet. This belief is based on decay of signal versus time that typically indicated approximately 1 to 2 hours for this system to "clean" up and approach the lower end of the range of mixing ratios stated above. The instrument's background was continuously monitored by blocking of the 1.1-μm laser beam at a frequency of 1 Hz and by periodically tuning the
226-nm laser's wavelength off of a NO resonant transition. Significant systematic changes were not detected in the numerous zero checks, calibrations, and converter efficiency tests of the NO\textsubscript{y} instrument, except for one noticeable reduction in the conversion efficiency for C\textsubscript{2}H\textsubscript{4}NO\textsubscript{2} (0.96 to 0.75). This decrease occurred after sampling a smelter plume during the ABLE 3B return transit flights on mission 21. Subsequent NO\textsubscript{y} measurements taken during the remainder of transit flights (missions 21 and 22) were reported as lower limits, and these data have not been used in this study.

The calibrations used to derive each flight's ambient NO, NO\textsubscript{a}, and NO\textsubscript{y} mixing ratios had, at the 95% confidence limit, standard deviations about the mean of ±16%, ±18%, and ±18%, respectively. The limit of detection (LOD) for a signal-to-noise ratio of 2/1 averaged 2 pptv for NO and 6 pptv for NO\textsubscript{y} using a 3-min signal integration time. The photon statistics based measurement precision at the 95% confidence limit (3 minutes integration time) was typically ±25% for NO = 15 pptv, ±35% for NO\textsubscript{y} = 35 pptv, and ±8.5% for NO\textsubscript{a} = 700 pptv. Measurement precision was proportional to the square root of mixing ratio and/or integration time.

The TP/LIF NO technique was used in both the NASA/GTE CITE 1 (1983) and the CITE 2 (1984) airborne intercomparison programs [Hoell et al., 1987; Gregory et al., 1990a]. These intercomparisons concluded that at low mixing ratios (i.e., < 60 pptv), NO measurements agreed with the stated instruments' precision and accuracy. The level of agreement among instruments was of the order of 35%. On average, individual measurements agreed within 5 to 7 pptv for NO mixing ratios of < 20 pptv. Concurrently, the photofragmentation TP/LIF NO\textsubscript{y} technique was evaluated during the CITE 2 airborne intercomparison program [Gregory et al., 1990b]. The NO\textsubscript{y} intercomparison study also concluded that, on average, the NO\textsubscript{y} measurements agreed within the stated instrument precision and accuracy at low NO\textsubscript{y} mixing ratios (i.e., NO\textsubscript{y} < 200 pptv). Individual measurements showed a 30 to 40% level of agreement (i.e., 15 to 20% from the average). Intercomparison at the lowest NO\textsubscript{y} mixing ratios (i.e., NO\textsubscript{y} < 50 pptv) revealed individual measurements were uncorrelated due to the randomness associated with approaching the instruments' limits of detection. On average, NO\textsubscript{y} values still agreed within the instruments' stated accuracy [Gregory et al., 1990a,b]. The instruments used in the ABLE 3A and ABLE 3B field programs were nearly identical in their evaluated precision and accuracy for NO and NO\textsubscript{y}.

2.2. Peroxyacetyl Nitrate (PAN), Peroxypropionyl Nitrate (PPN) Measurements

PAN and PPN were measured using a cryotrap (CT) preconcentration sample loop in conjunction with gas chromatographic (GC) separation, and a tandem volumetric electron capture detection (ECD) system. In the CT/GC/ECD technique, typically 0.10 to 0.20 standard liters of ambient air, drawn through an aft-facing Teflon inlet, were enriched in a cryotrap held at a constant -150°C temperature. The preconcentrated samples were then analyzed using a GC/ECD that was operated at constant pressure (1050 mbar).

Calibration of the PAN instrument was accomplished using diffusion tubes containing PAN standards that were prepared by the CH\textsubscript{3}CHO/NO\textsubscript{2}/Cl\textsubscript{2} photolysis method [Singh and Salas, 1983] and dissolved/stored in liquid n-tridecane [Gaffney et al., 1984]. Air was passed over the diffusion tube at a constant rate of flow to provide a calibration gas stream having PAN mixing ratios in the ppbv range. This calibration gas stream was periodically added at a downstream point in the ambient sampling line. PAN mixing ratios in the calibration gas stream were measured onboard the aircraft using a hot molybdenum oxide converter (375°C) coupled to a chemiluminescence NO monitor. Calibration of this chemiluminescence system was accomplished using both NO\textsubscript{2} and NO standards that were intercompared with NIST primary standards as part of the CITE 2 and ABLE 3B field program protocols. The calibration gas stream was further diluted using a two-stage serial dilution system to provide final PAN calibration mixing ratios in the parts per trillion by volume range. Linear mass flow controllers/meters used in this dilution system were compared with volume displacement standards. The accuracy of the PAN calibration transfer to ambient air measurements is estimated to be ±25% at the 95% confidence level.

The CT/GC/ECD PAN instrument was shown to be linear for PAN mixing ratios ranging from 5 to 1000 pptv. The typical limit of detection of the CT/GC/ECD PAN instrument was < 5 pptv for 100 ml of sampled air. Measurement precision was ±10% at the 95% confidence level for PAN mixing ratios that were well above the limit of detection (i.e., PAN ≥ 50 pptv).

This CT/GC/ECD PAN instrument was used in the NASA/GTE CITE 2 airborne intercomparison program [Gregory et al., 1990c]. This intercomparison program pointed out the need for onboard verification of PAN standards. As outlined above, onboard PAN standard verification was carried out during both ABLE 3A and ABLE 3B. The CITE 2 airborne intercomparison concluded that the two PAN instruments agreed, on the average, to about 20 pptv for PAN mixing ratios < 100 pptv. At larger PAN mixing ratios, agreement between individual measurements and the average was possible at the ±30% level for a 95% confidence limit even though individual pairs of measurements might sometimes fall outside this range. The PAN measurements that were intercompared fell close to agreeing within the stated accuracy and precision of the instruments [Gregory et al., 1990c].

2.3. HNO\textsubscript{3} Measurements

Gas phase HNO\textsubscript{3} was measured using a mist-chamber (MC) aqueous scrubber as a preconcentrator for subsequent determination of NO\textsubscript{y} by ion chromatography
ABLE 3A averaged 80% to 85% [Talbot et al., 1992a]. were corrected for a passing efficiency of 80%.

The HNO₃ mixing ratios reported for the ABLE-3A program were consistently at or below the LOD of the IC system prepared from dried KNO₃. These nitrate standards have a passing efficiency for HNO₃, and the uncertainties in the IC analysis of NO₃⁻, the results of laboratory tests of the MC's collection and the inlet's passing efficiencies for HNO₃, and the uncertainties in the measurement of the sampled air volume. The sampling manifolds and inlet system that drew in approximately 300 slpm of ambient air. All flow rates were measured with linear mass flow meters that were calibrated prior to and after each field program. A Teflon-coated inlet/sampling manifold was used during the ABLE 3A field program, whereas a 40-mm ID porcelain-glass-coated inlet/sampling manifold was used during the ABLE 3B program. This latter inlet/sampling manifold was nearly identical to the design used by the NO₃/NO₃⁻ instrument described previously. Laboratory tests of the porcelain-glass-coated inlet/manifold demonstrated 100% passing efficiency for HNO₃ mixing ratios in the 100-300 pptv range using flow rates of approximately 300 slpm. The change of inlet/manifold coating materials was prompted by concerns over possible sampling memory effects from a Teflon-coated system. HNO₃ passing efficiency of the 25-mm ID Teflon-coated inlet used in ABLE 3A averaged 80% to 85% [Talbot et al., 1992a]. HNO₃ mixing ratios reported for the ABLE-3A program were corrected for a passing efficiency of 80%.

The MC/IC calibrations were based on solutions prepared from dried KNO₃. These nitrate standards have agreed within ±3% of NIST standard solutions. The IC limit of detection (LOD) for HNO₃ was equivalent to 20 pptv for a 30-min sample collection time, using 30 slpm sample flow rates, where the LOD is inversely proportional to changes in the sample collection time. Blanks were obtained prior to each flight by sampling ambient air that was scrubbed from HNO₃ after passing through a series of impregnated and activated charcoal filters. Blank values were consistently at or below the LOD of the IC system using a 30-min sample collection time. Accuracy of the ICs calibration transfer to ambient HNO₃ measurements is estimated to be ±20% at the 95% confidence level, based on the uncertainties in the IC analysis of NO₃⁻, the results of laboratory tests of the MC's collection and the inlet's passing efficiencies for HNO₃, and the uncertainties in the measurement of the sampled air volume. The sampling time used during ABLE 3A ranged from 15 minutes to 3 hours, whereas those used during the ABLE-3B program ranged from 3 to 45 minutes.

The MC/IC technique has yet to participate in an airborne intercomparison study. The MC/IC technique has been used in a recent ground-based HNO₃ intercomparison. The HNO₃ measurement techniques used in this ground-based intercomparison included the MC/IC technique discussed here, a NOAA nylon filter collection technique, and a continuous flow liquid diffusion scrubber (preconcentrator) based system developed by Lind at NCAR. In all three techniques, ion chromatography was used to measure the NO₃⁻ that was collected by the preconcentrators. The preliminary results of this intercomparison indicated that on average the individual techniques agreed within ±25% of the mean value formed between pairs of measurements over the HNO₃ mixing range of 100 pptv to 500 pptv. Even so, individual pairs of HNO₃ measurements fell outside this range (E. L. Atlas et al., An intercomparison of three HNO₃ measurement techniques, submitted to Journal of Geophysical Research, 1992).

2.4. Particulate NO₃⁻ Measurements

Particulate NO₃⁻ (p-NO₃⁻) was measured using an isokinetic aerosol sampling probe with a curved-leading edge nozzle design [Talbot et al., 1992a,b]. A shrouded version of this nozzle design was used during the ABLE 3B field program. The collection efficiency of supermicron particles was greater than that for equivalent inlets using a sharp-leading edge nozzle design similar to those evaluated by Hubert et al. [1990]. Even so, the collection efficiency for supermicron particles must still be considered as uncertain. For submicron particles the curved leading edge nozzle has been shown to give significantly larger passing efficiency than straight edge designs with minimal loss on the nozzle inlet or curved tube sampling manifold wall (R. W. Talbot et al., Improvements in aerosol inlet performance in airborne applications, submitted to Journal of Atmospheric and Oceanographic Technology, 1992). The Teflon 90-mm filters were mounted on supports made of Delrin and contained by a Delrin housing.

Particles were collected on a stacked set of Nuclepore (8 µm) and Zefluor (1 µm) filters during the ABLE 3A program, whereas a single Zefluor (2 µm) filter arrangement was used in the ABLE 3B program. Sample collection times ranged from 15 minutes to 2 hours using a nominal ambient airflow rate of 475 lpm. These filter samples were analyzed for soluble NO₃⁻ by ion chromatography.

Filter blanks for p-NO₃⁻ were equivalent to ambient mixing ratios of 4 pptv and were subtracted from reported values. The accuracy of the ambient air equivalent NO₃⁻ that was measured on the filters was estimated to be ±20% at the 95% confidence level, excluding uncertainties involving the sample inlet passing/collection efficiency.

Some fraction of fine p-NO₃⁻ containing aerosols was most likely collected and converted in the NO₃⁻ system. This fraction is believed to be significantly smaller than that collected by the curved leading edge aerosol sampling system. Even though there is some degree of uncertainty on an absolute scale, the p-NO₃⁻ measurements that were made by the curved leading edge sampler described here should allow for an examination of relative trends in fine p-NO₃⁻, with respect to closure of the measured N₂O₅ family budget.
3. RESULTS AND DISCUSSION

3.1. Air Mass Characteristics

The Alaskan portion of the ABLE 3A program (missions 6-26) primarily overflew study regions near Barrow and Bethel, Alaska. Air masses encountered in the ABLE 3A program originated from regions that were typically free from anthropogenic sources based on 3- to 5-day air mass back-trajectory analyses. These air mass origins were determined to be from two primary source regions, namely, the Gulf of Alaska/Bering Sea, and northern Siberia/Arctic packice. Missions 14-21 were characterized by air masses originating from the Gulf of Alaska and Bering Sea, whereas missions 6-13 and 24-26 were characterized by air masses originating over northern Siberia and the Arctic packice [Shipham et al., 1992].

In general, the air masses sampled over Alaska had chemical signatures that suggested inputs of upper tropospheric to lower stratospheric air, biomass burning emissions from Alaska and Siberia, and longer-range transport of anthropogenic emissions from both Siberian and midlatitude Euro-Asian sources [Sandholm et al., 1992; Jacob et al., 1992; Wofsy et al., 1992].

The ABLE 3B program focused on two primary study areas, the Hudson Bay lowlands over northern Ontario/Manitoba (missions 2-9) and the boreal forested regions of northern Labrador and Quebec (missions 11-19). The air mass origins for these study areas, based on 3- to 5-day back-trajectory analyses, depicted a wide range of source regions that varied as the polar jet migrated between approximately 52° N and 66° N [Shipham et al., this issue]. The air masses sampled over the Hudson Bay lowlands generally originated from Alaska and the Northwest Territories. These air masses possessed chemical signatures that suggest inputs of upper tropospheric to lower stratospheric air and air influenced by biomass burning in western Canada and Alaska [Talbot et al., this issue; Anderson et al., this issue; Shipham et al., this issue].

Air masses sampled over the northern Quebec/Labrador region appeared to have a mixture of source influences. Biomass burning inputs were most notable on missions 14-19 possessing chemical signatures similar to what could be characterized as tropical air. These tropical air masses are believed to have been remnants of tropical Pacific air in association with Typhoon Steve and with remnants of tropical Atlantic air in association with Hurricane Bertha [Anderson et al., this issue; Blake et al., this issue; Shipham et al., this issue]. In addition to these tropically influenced air masses, air masses with characteristics of upper tropospheric to lower stratospheric origin were also encountered, most notably on missions 15-19 [Talbot et al., this issue; Browell et al., this issue].

In general, only the transit flight from Ontario to Labrador (mission 10) and a low-altitude flight leg over the Atlantic Ocean (mission 16) exhibited evidence of anthropogenic inputs from midlatitude sources in association with a strong southerly airflow.

3.2. NO₂ Measurements

The various NOₓ(i) measurements made during the ABLE 3 field programs covered a wide range of sampling time intervals. The HNO₃ (and p-NO₃⁻) measurement system used the longest sampling/preconcentration times with sampling times, which ranged from 3 min to 3 hours and covered spatial scales from about 25 km to 500 km. In general, the sampling times of this instrument were about 5 times longer during the ABLE 3A program than those used during the ABLE 3B program. PAN and PPN measurements were taken using 1- to 2-min sampling times with individual samples taken every 6- to 10-min. NOₓ and NO signals were recorded for 30-s signal integration times with individual measurements reported using longer integration times ranging from 1.5 to 3 min. Selection of the data set used here for investigating closure within the reactive nitrogen budget (i.e., NOₓ versus ΣNOₓ(i)) took into account this disparity in temporal overlaps of various NOₓ(i) and NOₓ measurements in relation to the magnitude of ambient variability observed in individual NOₓ(i) compounds during the period of comparison.

Because the HNO₃ measurement had the longest sample integration time, we chose to evaluate the budget based upon examining the average values of PAN, PPN, NOₓ, NOₓ, NOₓ, and other chemical variables observed during the HNO₃ sample integration period.

To account for effects that might arise from poor temporal overlap of the data during periods with large ambient variability, our NOₓ to ΣNOₓ(i) budget comparison data subset was filtered at two levels. First, the data were filtered to include only those HNO₃ measurement periods when all other NOₓ(i) compounds (PAN, PPN, and NOₓ) and NOₓ were also measured. This filter required a 65% temporal coverage by the NOₓ and NOₓ measurements and similar temporal coverage by PAN within the framework of the PAN instrument’s lower measurement duty cycle (i.e., PAN measurements 20% maximum temporal overlap could yield 13% temporal coverage of the HNO₃ measurement interval). The second filter that was used involved the examination of the temporal behavior of NOₓ, NOₓ, and PAN during individual HNO₃ measurement periods. Cases in which significant changes occurred in either NOₓ, NOₓ, or PAN mixing ratios that were not accompanied by measurements of all three compounds were excluded from the filtered data set.

Figures 1-3 display three examples of poor temporal overlap among the various NOₓ measurements during periods of ambient variability. In the first two cases, NOₓ mixing ratios were observed to increase by twofold to threefold in the middle of an individual HNO₃ measurement period, with PAN measurements occurring either outside of (see Figure 1), or at the center of (see Figure 2) the time period over which NOₓ mixing ratios peaked. Each figure also illustrates the fraction of HNO₃, NOₓ, and PAN (+ PPN) to NOₓ for the individual HNO₃ measurement period depicted in the time series plots and for nearby HNO₃ measurement periods that were selected by our filter to be valid NOₓ budget comparison data.
Fig. 1. Time series of different NO$_x$(i) species and NO$_y$ for Arctic Boundary Layer Expedition (ABLE) 3A mission 14, where mixing ratios (parts per trillion by volume) are represented by PAN (dash), HNO$_3$ (dash), NO$_x$ (cross), and NO$_y$ (asterisk), and the bar graphs for both the absolute mixing ratios and the fraction of NO$_y$ are in the bottom left and right, respectively.

points. Figure 3 displays an example in which the HNO$_3$ measurement period coincided with a large increase in both NO$_x$ and NO$_y$. In this case, PAN was measured only during the trailing edge of the event. All three of these cases and others like them were dropped from consideration due to poor temporal coverage during periods of ambient variability, as defined by a greater than 1.5-fold change in NO$_y$ mixing ratios over a HNO$_3$ measurement period without concurrent measurements of PAN and NO$_x$.

There were 84, 134, and 144 HNO$_3$ measurements made over the Alaska, Hudson Bay lowland, and Labrador boreal forest study regions, respectively. The filtering, discussed above, produced a budget comparison data subset that retained 34, 69, and 89 HNO$_3$ measurement periods, or 40%, 51%, and 62% of the original database. The larger retention of data from the ABLE 3B database was primarily due to a combination of shorter HNO$_3$ sampling times and a 1- to 1.5-hour transit time between the base airports and the intensive study areas. These transits generally consisted of stair-step flight profiles that were optimized for the study of N$_x$O$_y$ budgets in the free troposphere.

Table 1 gives a summary of the N$_x$O$_y$ composition observed in the filtered data set for the Alaska, the Hudson Bay lowlands, and the northern Labrador/Quebec study regions. These data have been separated into measurements made in the middle free troposphere (from 3 km to 6.2 km) and those made in the planetary boundary layer (taken here as < 3 km). Inspection of Table 1 reveals that mean NO$_x$ mixing ratios over both Canadian study regions were larger than those observed over Alaska by approximately 1.5-fold above 3 km and by 2-fold below 3 km. Over Canada the enhanced NO$_x$ mixing ratios below 3 km may reflect the combination of warmer temperatures and the approximate threefold larger PAN mixing ratios. This increase in both the average NO$_x$ and the PAN mixing ratios has been attributed to biomass burning [Talbot et al.,

Fig. 2. Time series of different NO$_x$(i) species and NO$_y$ for Arctic Boundary Layer Expedition (ABLE) 3A mission 17, where mixing ratios (parts per trillion by volume) are represented by PAN (dash), HNO$_3$ (dash), NO$_x$ (cross), and NO$_y$ (asterisk), and the bar graphs for both the absolute mixing ratios and the fraction of NO$_y$ are in the bottom left and right, respectively.
this issue]. Mean PAN mixing ratios above 3 km were nearly constant over all three study regions, even though median values were approximately 20% smaller over Alaska. Like its NO₃ precursor (and PAN reservoir), below 3 km mean HNO₃ mixing ratios were also larger by approximately 1.5-fold over Canada. However, above 3 km, mean HNO₃ mixing ratios were approximately twofold smaller over Canada, even though median values were similar for all three locations. This difference between mean and median values may reflect a lower frequency of precipitation for the air masses encountered over Alaska [Shipham et al., 1992; Shipman et al., this issue]. Overall, the ΣNO₂(i) above 3 km was relatively constant between all study regions, whereas below 3 km the values varied with the changes discussed above in PAN, NOₓ, and HNO₃. In contrast, below 3 km, NOₓ mixing ratios were approximately the same over Alaska and the Hudson Bay lowlands, whereas over northern Labrador/Quebec the values were approximately 35% smaller. Above 3 km, NOₓ mixing ratios changed by approximately a factor of 2 between the Hudson Bay lowlands and the northern Labrador/Quebec study regions, with the average value over Alaska falling in between.

3.3. Comparison of the Partitioning Observed in Nonurban Air

The partitioning of compounds within the N₂O₅ budget is summarized in Table 2. Also included are the partitionings observed in several other field programs that were recently summarized by Ridley [1991]. The comparison listed in Table 2 omits the contributions from p-NO₃⁺ due to the uncertainties associated with its contribution to measured NOₓ, or for that matter the “reactive” N₂O₅ budget (see also discussions by Atlas et al. [1992]). PAN constitutes the largest fraction of NOₓ in the Alaskan and Canadian middle free troposphere during the ABLE 3 study periods. The PAN fraction in this altitude range is similar to the middle free tropospheric values observed over the western continental United States during the CITE 2 field program even though the mean mixing ratio of PAN is approximately twofold smaller in the CITE 2 study. This smaller mixing ratio of PAN may reflect the warmer temperatures, which favor a faster thermal decomposition rate of PAN. This trend is maintained in the dramatically smaller PAN/NOₓ fraction and mixing ratios observed at Mauna Loa, Hawaii (cf. PAN lifetimes given by Singh et al. [1990] and Singh et al. [1992b]).

As with other measurements in the remote troposphere, NOₓ contributes only a small fraction to the NOₓ budget observed over the North American high latitudes. The NOₓ fraction of the NOₓ budget (NOₓ/NO₅) was smallest (~ 0.05) in the middle free troposphere over Alaska and the Hudson Bay lowlands. In these cases the median NOₓ/ΣNO₂(i) ratio (~0.12) is similar to values observed over northern Labrador/Quebec, the eastern Pacific, the western continental United States, and at Mauna Loa. In the northern Labrador/Quebec air masses that were characterized as having tropical origins, mixing ratios of NOₓ were within the range of values that have been reported for midtropospheric air masses over the Pacific (compare recent review by Ridley [1991] and Table 2 to Table 3). In these modified tropical air masses, mixing ratios of NOₓ also fall within the range of values for downslope (free tropospheric) flow conditions at Mauna Loa, even though PAN mixing ratios are considerably larger than those observed at Mauna Loa.

HNO₃ also represents a small fraction (0.12) in the high northern latitude middle free troposphere. In the 3- to 6.2-km altitude range over the Alaska and Hudson Bay lowland regions, the HNO₃ fraction of NOₓ is small in comparison to nonurban values from other studies, with the exception of those at Niwot Ridge, Colorado. These trends may be due to the variable scavenging and wash-out rates affecting different air masses and suggest HNO₃ is a poor choice to use for identifying trends in many cases.
TABLE 1. Summary of N₂O₅ Compounds Observed Over Alaska and Canada

| NO  | NO₂ | NOₓ | PPN | PAN | HNO₃ | p-NO₂ | ΣNO₂(i)NO₂ | Res. | Temp. | D. P. | CO   | O₃   | PAN/NO₂ | HNO₃/NOₓ |
|-----|-----|-----|-----|-----|------|-------|------------|------|-------|-------|------|------|--------|----------|
|     |     |     |     |     |      |       |            |      |       |       |      |      |        |          |
| Alaska, < 3km (N = 23) |
| Mean | 7.4 | 15  | 23  | NR  | 35   | 73    | 28         | 125  | 349   | 223   | 9.6  | 4.1  | 103    | 37       |
| Median| 6.6 | 14  | 19  | NR  | 17   | 50    | 16         | 90   | 284   | 194   | 11   | 5.8  | 103    | 32       |
| Standard Deviation| 4.3 | 6.3 | 9.8 | NR  | 43   | 62    | 24         | 104  | 186   | 120   | 5.5  | 5.2  | 15     | 11       |
| Minimum | 3.6 | 7.5 | 15  | NR  | 7.5  | 20    | 7          | 47   | 157   | 30    | -2.6 | -8.3 | 86     | 21       |
| Maximum | 22  | 33  | 55  | NR  | 209  | 335   | 103        | 572  | 1010  | 437   | 19   | 11   | 141    | 70       |
| Alaska, > 3km (N = 11) |
| Mean | 11  | 19  | 31  | 9   | 278  | 100   | 26         | 397  | 669   | 272   | -11  | -23  | 113    | 78       |
| Median| 11  | 19  | 31  | 9   | 234  | 45    | 17         | 378  | 639   | 305   | -12  | -24  | 112    | 78       |
| Standard Deviation| 3.5 | 4.4 | 6.6 | 1.7 | 89   | 101   | 23         | 175  | 185   | 74    | 4.4  | 3.9  | 25     | 2.7      |
| Minimum | 5.5 | 11  | 18  | 7   | 162  | 20    | 12         | 207  | 434   | 126   | -18  | -32  | 96     | 66       |
| Maximum | 18  | 26  | 39  | 12  | 458  | 375   | 96         | 840  | 1180  | 347   | -4.8 | -18  | 138    | 90       |
| Hudson Bay Lowlands, < 3km (N = 43) |
| Mean | 10  | 38  | 48  | 0.2 | 95   | 120   | 11         | 252  | 340   | 89    | 16   | 7.7  | 127    | 37       |
| Median| 8.5 | 32  | 41  | 0   | 50   | 81    | 7.8        | 190  | 251   | 102   | 16   | 7.3  | 132    | 33       |
| Standard Deviation| 4   | 21  | 23  | 0.6 | 89   | 101   | 180        | 195  | 65    | 4.8   | 3.4  | 10   | 20     | 10       |
| Minimum | 5   | 15  | 21  | 0   | 13   | 13    | 4          | 62   | 112   | -38   | 5.8  | 2.1  | 94     | 25       |
| Maximum | 22  | 95  | 103 | 3   | 318  | 338   | 40         | 670  | 749   | 268   | 28   | 14   | 160    | 73       |
| Hudson Bay Lowlands, > 3km (N = 26) |
| Mean | 13  | 35  | 48  | 0.7 | 302  | 67    | 30         | 399  | 841   | 442   | -9.4 | -21  | 109    | 64       |
| Median| 11  | 30  | 43  | 0   | 317  | 51    | 8.2        | 386  | 735   | 345   | -11  | -21  | 99     | 61       |
| Standard Deviation| 6.5 | 15  | 21  | 1.1 | 148  | 49    | 54         | 144  | 400   | 358   | 5.8  | 12   | 22     | 13       |
| Minimum | 4.7 | 16  | 21  | 0   | 41   | 10    | 4          | 191  | 383   | 124   | -18  | -42  | 88     | 45       |
| Maximum | 28  | 88  | 116 | 5   | 756  | 212   | 180        | 849  | 2000  | 1540  | 3.0  | 3    | 176    | 93       |
| Northern Labrador/Quebec, < 3km (N = 52) |
| Mean | 8   | 32  | 40  | 0.2 | 87   | 116   | 9.6        | 233  | 253   | 20    | 11   | 3.3  | 102    | 32       |
| Median| 6.9 | 25  | 33  | 0   | 61   | 69    | 7.1        | 191  | 202   | -3.5  | 10   | 2.5  | 96     | 28       |
| Standard Deviation| 4   | 24  | 28  | 0.9 | 65   | 106   | 7.6        | 164  | 190   | 88    | 6.7  | 7.3  | 20     | 12       |
| Minimum | 3.6 | 12  | 18  | 0   | 14   | 11    | 4          | 52   | 60    | -177  | -1.7 | -22  | 82     | 18       |
| Maximum | 29  | 142 | 158 | 4   | 280  | 437   | 40         | 807  | 920   | 201   | 24   | 17   | 166    | 59       |
| Northern Labrador/Quebec, > 3km (N = 37) |
| Mean | 9.8 | 33  | 43  | 0.6 | 295  | 50    | 6.5        | 371  | 459   | 89    | -7.9 | -18  | 100    | 55       |
| Median| 8.4 | 30  | 45  | 0   | 298  | 45    | 4          | 394  | 338   | -14   | -6.4 | -20  | 100    | 57       |
| Standard Deviation| 4.5 | 12  | 14  | 1.5 | 106  | 36    | 4.2        | 124  | 487   | 444   | 5.9  | 9    | 18     | 8.6      |
| Minimum | 3.1 | 15  | 20  | 0   | 49   | 10    | 3.7        | 94   | 113   | -175  | -21  | -41  | 64     | 35       |
| Maximum | 18  | 72  | 75  | 6   | 518  | 136   | 19         | 621  | 3250  | 2680  | 2.0  | 0.3  | 142    | 79       |

ΣNO₂(i) does not include p-NO₂. Mean and median ΣNO₂(i) residual (Res.), PAN/NO₂, and HNO₃/NOₓ are taken from individual data points (e.g. not mean of means). Mixing Ratios of NO, NOₓ, PAN, HNO₃, p-NO₂, ΣNO₂(i), NO₃, and residual (Res. = NO₂-ΣNO₂(i)) are in parts per trillion by volume. Mixing Ratios of CO and O₃ are in parts per billion per volume, temperature (°C), D.P. (Dew Point °C).
TABLE 2. Partitioning and Balance of $N_xO_y$

|                | $NO_y/NO_x$ | $HNO_3/NO_x$ | $PAN/NO_x$ | $\Sigma NO_y(i)/NO_x$ | $NO_y$ | $CO/NO_x$ | $O_3/NO_y$ | $C_2H_4/C_2H_6$ |
|----------------|-------------|--------------|------------|-----------------------|--------|----------|-----------|----------------|
| **Alaska, 1988** |             |              |            |                       |        |          |           |                |
| < 3 km         | 0.07        | 0.18         | 0.07       | 0.32                  | 275    | 0.37     | 0.12      | 0.12          |
| 3 - 6.2 km     | 0.045       | 0.08         | 0.39       | 0.52                  | 650    | 0.17     | 0.12      | 0.12          |
| **Hudson Bay Lowlands, 1990** |             |              |            |                       |        |          |           |                |
| < 3 km         | 0.15        | 0.34         | 0.19       | 0.73                  | 250    | 0.24     | 0.12      | 0.1           |
| 3 - 6.2 km     | 0.055       | 0.065        | 0.4        | 0.51                  | 750    | 0.14     | 0.085     | 0.11          |
| **Northern Labrador/Quebec, 1990** |             |              |            |                       |        |          |           |                |
| < 3 km         | 0.19        | 0.37         | 0.39       | 1.03                  | 200    | 0.73     | 0.16      | 0.11          |
| 3 - 6.2 km     | 0.12        | 0.095        | 0.79       | 1.01                  | 350    | 0.27     | 0.16      | 0.11          |
| **Modified Tropical Air Over Northern Labrador/Quebec, 1990** |             |              |            |                       |        |          |           |                |
| 3 - 6.2 km     | 0.16        | 0.09         | 0.64       | 0.83                  | 225    | 0.37     | 0.21      | 0.035         |
| **Eastern Pacific, Summer 1986 \* CITE 2** |             |              |            |                       |        |          |           |                |
| 4.6 - 5 km     | 0.09        | 0.32F-0.49D  | 0.33       | 0.78F-0.9D            | 425    |          |           |                |
| 4.6-6.1 km     | 0.14        | 0.32F-0.16D  | 0.42       | 0.9F-0.7D             | 375    |          |           |                |
| **Niwot Ridge, Summer 1984 \†** |             |              |            |                       |        |          |           |                |
| 3 km           | 0.25        | 0.095        | 0.17       | 0.52                  | 2100   |          |           |                |
| **Niwot Ridge, Summer 1987 \*†** |             |              |            |                       |        |          |           |                |
| 3 km           | 0.32        | 0.10         | 0.24       | 0.73                  | 1600   |          |           |                |
| **MLOPEX Downslope, Spring 1988 \‡** |             |              |            |                       |        |          |           |                |
| 3.4 km         | 0.14        | 0.43         | 0.047      | 0.79                  | 253    |          |           |                |

The values in the table are median ratios not ratios of medians unless otherwise noted. Ratios of $NO_x$, $HNO_3$, $PAN$, and $\Sigma NO_y(i)$ to $NO_y$ and $C_2H_4/C_2H_6$ are pptv/pptv, whereas CO and $O_3$ to $NO_y$ are ppbv/pptv.

\*From Ridley [1991], where the range of values given in Chemical Instrumentation Test and Evaluation 2 denote the range of $HNO_3$ mixing ratios using two different techniques and where (F) and (D) denote filter pack or denuder $HNO_3$ measurements.

\†From Fahey et al. [1986], where the values are ratios of means extrapolated from their Figure 12 for daytime measurements.

\‡From Atlas et al. [1992], where $\Sigma NO_y(i)/NO_y$ includes contributions from $p-NO_3$ ($-0.06$) and alkyl nitrates ($-0.01$).

In the middle free troposphere over Alaska and Hudson Bay lowlands, only about 55% of the measured $NO_y$ was accounted for by the measured $NO_y(i)$ species, which translates to a budget deficit ranging from approximately 300 to 400 pptv. At lower altitudes the $\Sigma NO_y(i)$ accounted for approximately 40 to 70% of the measured $NO_y$, resulting in $NO_y$ budget deficits ranging from approximately 200 to 100 pptv, respectively. Ridley [1991] has presented arguments against in situ production increasing the absolute abundance of unaccounted for $NO_y(i)$ compounds during air mass aging. The relative abundance of chemically stable unaccounted for $NO_y(i)$ compound or compounds could, however, be increased during air mass aging as the active $NO_x$ pool is converted via OH oxidation (or $N_2O_5$ hydrolysis) to $HNO_3$, which can be lost via scavenging processes. The magnitude of the deficit over Alaska and the Hudson Bay lowlands is larger, by about twofold to fourfold, than lower-latitude free tropospheric observations. The magnitude of the deficit in the $NO_y$ budget is, however, at the lower end of the 0.2 to 1.5 ppbv range of mixing ratios that would occur from the 6 to 25% $NO_y$ deficits that have been observed in rural areas of the continental United States [e.g., Parrish et al., 1993]. The influx and dilution of these quantities of unaccounted for $NO_y$ compounds could contribute to a portion of the $NO_y$ budget deficit observed over some high-latitude regions [Sandholm et al., 1992; Wofsy et al., 1992].

For industrial pollutants to represent the compounds that form the deficits observed at all altitudes, they would, however, need to be more thermally stable than PAN. As previously discussed, there were some indications that the decomposition of PAN could not fully account for the $NO_y$ mixing ratios observed in the 4- to 6-km altitude range over Alaska [Jacob et al., 1992]. If true, this would suggest a portion of the $NO_y$ deficit may act as a reservoir for the production of active odd nitrogen ($NO_y$). In the following sections we will attempt to address these hypotheses by investigating the factors that might be influencing the observed degree of closure within the high-latitude $N_xO_y$ budget.

3.4. Degree of Closure Within $N_xO_y$ Budget

In our examination of the degree of closure between $NO_y$ and the $\Sigma NO_y(i)$ within the filtered $NO_y$ budget comparison data subset, we feel it is necessary to take into account the measurement uncertainty of the individual $NO_y(i)$ species relative to $NO_y$. In general, each technique's claimed precision and accuracy has been generally demonstrated, based on the average observed mixing ratios using multiple techniques for the measurement of $NO_y$, $NO_x$, $PAN$, and $HNO_3$ (see earlier discussions in section 2). Based on the average level of agreement found in these intercomparison studies, the overall uncertainty at the 95% confidence limit for each measurement has been taken as $\pm 25\%$ for $NO_y$, $\pm 30\%$ for PAN (+ PPN), $HNO_3$, and $NO_y$. The relatively large uncertainty for $NO_y$ is due to our inability to perform an a priori assessment of
potential NO\textsubscript{y} sampling problems, as many different compounds with different chemical properties are potentially involved (see discussion later in section 3.5).

Figures 4 and 5 display the degree of balance in the N\textsubscript{Oy} budget based on calculating the residual = NO\textsubscript{y} - \Sigma NO\textsubscript{y}(i) (Figure 4) and the relative degree of balance (i.e., residual/NO\textsubscript{y}, Figure 5) for the entire filtered N\textsubscript{Oy} budget data subset, where the \Sigma NO\textsubscript{y}(i) consists of NO\textsubscript{y}, PAN (+PPN), and H\textsubscript{NO}. Measurements of CH\textsubscript{3}ONO\textsubscript{2} were also reported for the Alaska studies. These CH\textsubscript{3}ONO\textsubscript{2} measurements have been excluded from this analysis because the values were usually small in comparison to those of PPN and nearly negligible in comparison to those of PAN. In addition, the residual was calculated assuming an NO\textsubscript{y} conversion efficiency of 0.95 for both PAN (+PPN) and H\textsubscript{NO}, based on the convertor efficiency characterizations results of Fahey et al. [1985]. These data have been segregated into three categories based on whether the absolute magnitude of the residual fell: (1) within one standard deviation of its estimated uncertainty (i.e., |residual| < 1σ, Figures 4a and 5a); (2) between one and two standard deviations of its estimated uncertainty (i.e., 1σ < |residual| < 2σ, Figures 4b and 5b); or (3) greater than two standard deviations of its estimated uncertainty (i.e., |residual| > 2σ, Figures 4c and 5c). This assessment was based on taking the overall uncertainty as the square root of the sum of the variances estimated above for NO\textsubscript{y}, PAN, PPN, H\textsubscript{NO}, and NO\textsubscript{y} (i.e., assuming that the variances are uncorrelated and random and the variance equals the square of the estimated measurement uncertainties).

Little altitude dependence is exhibited in either the absolute or the relative values of the residuals for the data having |residuals| < 1σ (see Figures 4a and 5a). In addition, these data are nearly evenly distributed about zero and are within the range ±100 pptv and ±0.2, respectively. The portion of the data having |residuals| between 1σ and 2σ (see figures 4b and 5b) is also distributed about zero, but with a larger number of points having positive residuals. In this category (i.e., 1σ < |residuals| < 2σ) the northern Labrador/Quebec data have fractional values that are more evenly distributed about zero than data from the Hudson Bay lowlands or Alaska. This tendency is especially pronounced for the portion of the data having |residuals| > 2σ. In the latter case, the number of points with negative residuals is not outside the range expected at the 95% confidence limit and tends to support the magnitudes of our overall uncertainty estimates. Based on these uncertainty constraints, we have adopted the following definitions for describing the NO\textsubscript{y} to \Sigma NO\textsubscript{y}(i) budget, where (1) out-of-balance characterizes those data with positive valued residuals > 2σ and (2) in-balance characterizes those data with |residuals| < 2σ.

The abundance of NO\textsubscript{y}(i) and fractional abundance (NO\textsubscript{y}(i)/NO\textsubscript{y}) are given in Figures 6-8 for the three study regions. The data in these figures have been sorted by the fractional magnitude of the residuals. In addition to the \Sigma NO\textsubscript{y}(i) and NO\textsubscript{y} data presented in Figures 6-8,
IResidual < 1σ

1σ < |Residual| < 2σ

|Residual| > 2σ

Fig. 4. Altitudinal dependence of the residual (residual = NOy - ΣNOy(j)) separated into one of three classifications: (a) |residual| is less than one standard deviation of the measurement uncertainties; (b) |residual| falls between one and two standard deviations of the measurement uncertainties; and (c) |residual| is greater than two standard deviations of the measurement uncertainties. A/a denote data from Alaska during ABLE 3A, L/l denote data from northern Labrador and Quebec during ABLE 3B, and H/h denote data from the Hudson Bay lowlands during ABLE 3B, where the uppercase letters indicate data taken between 3 and 6.2 km and lowercase letters indicate data taken at < 3 km.

particulate-nitrate (p-NO3') mixing ratios are given when available, but they were not used in the calculation of residual values. Inspection of Figures 6-8 reveals that in general, even if it were sampled efficiently by the NOy instrument, p-NO3' should have contributed negligibly to the NOy budget.

Below 3 km the NOy to ΣNOy(j) budget for the Alaska data is clearly out-of-balance, having on average only about 35% of the NOy accounted for by the ΣNOy(j). Above 3 km the data are also predominantly out-of-balance but with a larger fraction (~55%) of the NOy accounted for by the ΣNOy(j). Above 3 km the Hudson Bay lowland data are also more out-of-balance. However, below 3 km the Hudson Bay lowland data are more evenly distributed with respect to overall estimates of the residuals' uncertainty. For this case, NOy mixing ratios range from approximately 100 to 750 pptv, with a noticeable decrease in average values between the in- and out-of-balance categories (~400 pptv versus ~200 pptv, respectively). In contrast, over the northern Labrador/Quebec study region the data from both altitude ranges (i.e., > 3 km and < 3 km cases) are predominantly in-balance with the ΣNOy(j)/NOy ratio having a mean value near unity.

For comparison, a statistical summary of several of the in- and out-of-balance cases discussed above is given in Table 4. Below 3 km the out-of-balance Hudson Bay lowland data have mean mixing ratios that are approximately fourfold smaller for PAN and HNO3 and approximately twofold smaller for NO and NOy compared to the corresponding in-balance data. In this out-of-balance case the mean mixing ratios of the NOy(j) compounds and NOy are closer to those observed in the lower-altitude Alaska data, which are predominantly out-of-balance (if the smaller mixing ratios of HNO3 observed in the Hudson Bay lowland case were due to increased wet deposition). The in-balance northern Labrador/Quebec data have nearly identical median mixing ratios of NOy both above and below 3 km even though, as might be expected, PAN mixing ratios increased by approximately threefold in the higher-altitude regime. This increase in PAN mixing ratios more than offsets, by about twofold, the decrease in median HNO3 mixing ratios allowing PAN to be the key species controlling the NOy budget within this altitude range over northern Labrador/Quebec.

The larger fraction of in-balance data over northern Labrador/Quebec data may be due to a combination of factors. The dominant source of enhanced N2O5, mixing ratios into the middle to lower (i.e., < 6 km) troposphere over the northern Labrador/Quebec region has been attributed to relatively fresh biomass burning emissions [Talbot et al., this issue]. In addition, during the study period, there were several large influxes of what has been characterized as modified tropical air into the northern Quebec/Labrador region [see Talbot et al., this issue; Shipham et al., this issue]. In these air masses the mean mixing ratios of PAN, HNO3, and NOy are approximately twofold smaller than the corresponding mean values of the in-balance data set for this altitude range (cf. Tables 3 and 4). This suggests that the air masses over this region appear to have been influenced by both the influx of
Fig. 5. Altitudinal dependence of the residual (NOy - \( \Sigma \)NOy(i)) relative to the measured NOy mixing ratio (residual/NOy) separated into one of the three classifications: (a) |residual| is less than one standard deviation of the estimated uncertainties; (b) |residual| falls between one and two standard deviations of the estimated uncertainties; and (c) |residual| is greater than two standard deviations of the estimated uncertainties. The data labels are described in Figure 4.

3.5. NOy Budget Implications of Trends Between NOy and Other Components

Implications of the correlative trends between individual NOy(i) and other measured compounds to the sources and sinks of reactive nitrogen are discussed in more detail by Sandholm et al. [1992], Singh et al. [1992a, b, this issue], Wofsy et al. [this issue], and Talbot et al. [this issue]. Therefore in this section we will focus on investigating trends related to the NOy versus \( \Sigma \)NOy(i) budget issue. Figures 9a and 9b illustrate the average correlation between mixing ratios of the individually measured NOy(i) compounds (i.e., NOx, PAN, and HNO3) and NOv for both the in- and the out-of-balance data. Based on the linear regressions, these data suggest that on average there is an approximately threefold larger co-variance of NOx with respect to NOy for the in-balance data subset relative to the out-of-balance data subset. The small subset of out-of-balance data represented by a square is an aggregate of measurements made in plumes that had been cloud pumped to the aircraft's altitude (4 to 6 km; see also Figures 4 and 5). These plumes are believed to have originated from nearby biomass burning in the region. These data were not included in any of the calculated regressions for the out-of-balance data and, in general, the plume data will be discussed separately.

In contrast to the change in covariance implied from the NOx versus NOy regressions, the regression slopes for PAN versus NOy imply that on average there is little change in the covariance relationship between the in- and the out-of-balance data. This might suggest that on average the out-of-balance data may represent more aged air parcels that had experienced a larger loss of NOx (e.g., via OH + NOx \( \rightarrow \) HNO3) relative to PAN. The lack of an enhanced HNO3 to NOy relationship or for that matter the lack of any strongly perceivable relationship in the out-of-balance data appears to contradict this hypothesis. However, as indicated by the small HNO3 mixing ratios in the cloud-processed plume data, HNO3 solubility precludes it from being a good surrogate for inspecting covariance relationships and suggests the need to examine correlations with respect to other trace gases.

Significant correlations between mixing ratios of O3 and NOy have been reported in several other free tropospheric studies. These studies have yielded fairly consistent linear regression slopes that are on average near 0.1 ppbv O3/ppbv NOy. Of their reported studies,
only the AASE 2 data appeared to maintain a linear $O_3$ to $NO_y$ relationship at $NO_y$ mixing ratios larger than approximately 450 pptv. For $NO_y$ mixing ratios > 450 pptv, both the Mauna Loa Photochemistry Experiment (MLOPEX) and CITE 2 data exhibit a tendency toward less correlation between $O_3$ and $NO_y$, which is similar to the trend described by Sandholm et al. [1992] for composites of the ABLE 3A data taken over Alaska. The overall ABLE 3 $N_2O_5$ budget subset of data exhibits a generally similar behavior (see Figure 10a). The correlation between mixing ratios of $O_3$ and $NO_y$ appears to be a persistent feature in a significant portion of the northern hemisphere's free troposphere. In particular, the in- and out-of-balance data both exhibit similar degrees of correlation between $O_3$ and $NO_y$ (cf. Figures 10b and 10c), even though the slope for the out-of-balance data is somewhat smaller (i.e., ~0.06 versus ~0.1). As might be expected, the slope of the $O_3$ versus $\Sigma NO_y$ regression for the $NO_y$ < 450 pptv out-of-balance data is larger than the values obtained from the various $O_3$ versus $NO_y$ regressions (i.e., 0.13 versus approximately 0.1 to 0.06; see Figure 11a; [Hübner et al., 1992; Sandholm et al., 1992]. Surprisingly, the slope of the $O_3$ versus residual for the out-of-balance data is also in the range of values discussed above (see Figure 11b). This slope, in combination with the slope from $O_3$ versus $\Sigma NO_y(i)$, is of the correct magnitude to explain the smaller slope of the out-of-balance $O_3$ versus $NO_y$ regressions. This correlation between $O_3$ and residual values suggests that on average the residuals have some degree of covariance with the factors influencing the troposphere's oxidative potential, as portrayed by $O_3$ mixing ratios, over these high-latitude regions.

Correlations have also been observed between mixing ratios of $CO$ and $NO_y$ in midlatitude urban and nonurban air masses [Parrish et al., 1991]. These midlatitude data exhibited a nearly linear regression for mixing ratios of $NO_y$ > 1 ppbv, with a linear $CO$ versus $NO_y$ regression slope near 20 (ppbv/ppbv). Below 1 ppbv, little correlation between mixing ratios of $CO$ and $NO_y$ was observed in their midlatitude data or in the midlatitude free tropospheric measurements made at Mauna Loa, where $CO$ mixing ratios ranged from approximately 120 to 160...
ppbv [Hübner et al., 1992]. In contrast, the higher-latitude measurements of CO and NOy presented here exhibit a linear correlation for NOy mixing ratios < 1 ppbv and for CO mixing ratios generally below those found at Mauna Loa (see Figure 12a). The slope of this regression has a value that is approximately twofold larger than that obtained from the NOy > 1 ppbv midlatitude data presented by Parish et al. [1991]. The larger CO versus NOy slope may reflect the inputs of CO-rich/NOy-poor emissions from high-latitude smoldering fires. Ratios of NOy/CO have been observed to vary from approximately 0.0006 (±0.0003) for smoldering biomass burning [Wofsy et al., 1992] to approximately 0.026 (± 0.005) for moderately aged plumes originating from the eastern U.S. [Wofsy et al., this issue]. Based on laboratory experiments, biomass burning can have highly variable NOy/CO emission ratios depending on whether flaming or smolderous combustion conditions are being supported (NOy/CO ratios from about 1 to < 0.01, Lobert et al. [1990]). These variations in NOy/CO ratios may help explain the loss in correlation going from the in-balance to out-of-balance subsets of data ($r^2 = 0.92$ versus $r^2 = 0.70$, cf. Figures 12b and 12c). In conjunction, the relatively small ratios of CO/NOy in the out-of-balance cloud-pumped plume data suggests these plumes may have originated from sources such as flaming biomass burning or industrial/urban pollution. These combined results suggest that all of these chemical signatures have been influenced by the emissions from a variety of sources.

To study further the possible causes for these trends, we have also investigated trends between values of NOy, ΣNOy(i) and the residuals versus several surrogate time markers of the relative age of an air mass. We found little evidence of correlations with values of the parameter NOy−NO, which has been successfully used as a "chemical-clock" that reflects the degree of oxidation in urban plumes when NO is the dominant NOy(i) species. Our data's lack of correlation with this parameter, NOy−NO, is in agreement with the MLOPEX results presented by Atlas et al. [1992] and likely reflects their argument that in more aged air masses the NOy to NO "chemical clock" has been continually reset by loss processes (e.g., wet removal) and recycling of odd nitrogen species.

Ratios of various carbon-containing compounds have also been used in numerous studies as surrogate "chemical clocks" that are not prone to resetting by such factors as rain/washout. However, they can be reset by the mixing in of fresh emissions that can have a variety of source emission factors and by the dilution of air parcels through the mixing in of background air having various ages (see S. McKeen and S. C. Liu, Hydrocarbon ratios and photochemical history of air masses, submitted to Geophysical Research Letters, 1993). Thus mixing and photochemical aging processes both contribute to
establishing the ensemble average of an air mass's relative age. This precludes the unambiguous identification of sources and sinks based on any one "chemical clock's" estimate of an air mass's age and certainly precludes directly determining causation from any individual correlative trend. Thus, whereas it is tempting to argue that the on average correlation between \( \mathrm{O}_3 \) and \( \mathrm{NO}_y \) is suggestive of a common stratospheric source, it is perhaps more appropriate to suggest that this correlation merely reflects the common behavior of two species that are reasonably long-lived in the upper troposphere (as is \( \mathrm{CO} \)) and that share similar loss rates due to photochemical and/or deposition processes as air parcels become distributed (and diluted) on a hemispheric scale. The small HNO\(_3\) fraction of \( \mathrm{NO}_y \) in these data sets also support this latter argument, unless an efficient mechanism can be identified that is capable of converting stratospherically derived HNO\(_3\) back into more active forms of \( \mathrm{NO}_y \).

Recent laboratory measurements indicate that \( \mathrm{NO}_x \) can be generated from the surface catalyzed photolysis of \( \mathrm{H}_2\mathrm{SO}_4 \) and HNO\(_3\) mixtures [D. Fahey, private communication]. These observations could suggest that a similar aerosol coupled mechanism might be responsible for liberating the reactive nitrogen pool that has been thought to become irreversibly tied up as HNO\(_3\). The implications of such possible mechanism to the distribution of \( \mathrm{NO}_x \) over remote regions certainly warrant careful laboratory study. Another possible explanation is the residence time in the upper troposphere (> 6 km) is sufficiently long to allow OH oxidation and photolysis to slowly reanimate the stratospheric \( \mathrm{NO}_x \) pool that is tied up in HNO\(_3\). Ratios of carbon containing compounds can still, however, provide some useful information about a particular air mass's combined photochemical history relative to other air masses. In particular, the ratios \( \mathrm{C}_2\mathrm{H}_2/\mathrm{CO} \) and \( \mathrm{C}_3\mathrm{H}_8/\mathrm{C}_2\mathrm{H}_6 \) have been used to reflect the degree of combined mixing and photochemical processing that has occurred within an air mass (see discussion by McKeen and Liu [ibid]). During this atmospheric processing, the dominant tropospheric chemical loss process for all four compounds is their oxidative reactions with OH. The faster reaction rate coefficients [e.g., DeMore et al., 1992] for \( \mathrm{C}_2\mathrm{H}_2 \) and \( \mathrm{C}_3\mathrm{H}_8 \) versus \( \mathrm{CO} \) and \( \mathrm{C}_2\mathrm{H}_6 \) yield expectations that these two pair of ratios should have similar values that decrease as an isolated air mass photochemically ages. In our ABLE 3A investigations, \( \mathrm{NO}_y \) was well correlated with ratios of \( \mathrm{C}_2\mathrm{H}_2/\mathrm{CO} \) [Sandholm et al., 1992]. This tendency also generally holds for both the in- and out-of-balance data (see Figures 13a and 13b). However, the out-of-balance cloud-pumped plume data do not follow the trend expected from either the in-balance or the nonplume out-of-balance data. These plumes did appear to be depleted of HNO\(_3\) relative to \( \mathrm{NO}_x \) or PAN, as discussed earlier. Somewhat surprisingly though, these plumes tend to follow the general trend exhibited between the ratios of \( \mathrm{C}_2\mathrm{H}_2/\mathrm{CO} \) versus the
### TABLE 4. Summary of in- and out-of-Balance Case Studies

|        | NO  | NO₂ | NO₃ | PPN | PAN | HNO₃ | p-NO₂ | HNO₃(p-NO₂) | ΣNOₓ(i) | NOₓ | Res. | Temp. | NOₓ/NO₂ | PAN/NO₂ | HNO₃/NO₂ | ΣNOₓ(i)/NOₓ |
|--------|-----|-----|-----|-----|-----|------|------|------------|--------|-----|------|-------|---------|---------|----------|-------------|
| **Hudson Bay Lowlands, < 3km, in Balance** (N=19) |
| Mean   | 11  | 48  | 59  | 0.4 | 157 | 188  | 15   | 387        | 445    | 58  | 16   | 0.15  | 0.32    | 0.48    | 0.91     |
| Median | 9.7 | 46  | 58  | 0   | 171 | 194  | 12   | 429        | 532    | 70  | 16   | 0.15  | 0.37    | 0.41    | 0.87     |
| SD     | 4.3 | 20  | 22  | 0.9 | 99  | 69   | 13   | 155        | 201    | 62  | 5    | 0.05  | 0.12    | 0.17    | 0.12     |
| Minimum| 5.5 | 20  | 26  | 0   | 13  | 75   | 4    | 122        | 113    | -37 | 5.8  | 0.07  | 0.08    | 0.26    | 0.76     |
| Maximum| 22  | 92  | 100 | 3   | 318 | 330  | 40   | 669        | 750    | 154 | 27   | 0.23  | 0.45    | 0.8     | 1.2      |
| **Hudson Bay Lowlands, < 3km, out of Balance** (N=12) |
| Mean   | 9   | 23  | 32  | 0   | 36  | 28   | 6    | 92         | 237    | 145 | 15   | 0.13  | 0.15    | 0.12    | 0.39     |
| Median | 7   | 17  | 25  | 0   | 31  | 22   | 4.6  | 77         | 205    | 125 | 13   | 0.13  | 0.15    | 0.11    | 0.38     |
| SD     | 4.3 | 8.4 | 11  | 0   | 17  | 11   | 1.7  | 33         | 74     | 43  | 4.2  | 0.02  | 0.03    | 0.03    | 0.4      |
| Minimum| 6   | 15  | 22  | 0   | 22  | 13   | 4    | 62         | 179    | 107 | 9.5  | 0.1   | 0.1     | 0.07    | 0.28     |
| Maximum| 20  | 40  | 60  | 0   | 86  | 52   | 7.8  | 190        | 458    | 268 | 22   | 0.18  | 0.19    | 0.18    | 0.45     |
| **Northern Labrador/Quebec, < 3km, in Balance** (N=31) |
| Mean   | 8   | 33  | 41  | 0.2 | 90  | 111  | 9    | 232        | 237    | 4.7 | 11   | 0.21  | 0.43    | 0.45    | 1.03     |
| Median | 6.8 | 24  | 31  | 0.6 | 60  | 62   | 7.1  | 156        | 138    | -4.4| 9    | 0.21  | 0.42    | 0.4     | 1.04     |
| SD     | 4.9 | 28  | 32  | 0.8 | 67  | 110  | 6.4  | 180        | 214    | 54  | 7    | 0.07  | 0.15    | 0.19    | 0.15     |
| Minimum| 3.6 | 12  | 18  | 0   | 25  | 14   | 4    | 60         | 60     | -78 | -1.7 | 0.08  | 0.17    | 0.15    | 0.77     |
| Maximum| 29  | 142 | 158 | 4   | 280 | 437  | 23   | 808        | 920    | 162 | 24   | 0.38  | 0.89    | 1.04    | 1.3      |
| **Northern Labrador/Quebec, > 3km, in Balance** (N=26) |
| Mean   | 8.5 | 32  | 40  | 0.4 | 300 | 50   | 6.1  | 374        | 352    | -22 | -8.6 | 0.12  | 0.85    | 0.15    | 1.1      |
| Median | 6.6 | 29  | 35  | 0   | 298 | 39   | 4    | 392        | 332    | -24 | -7.9 | 0.12  | 0.84    | 0.11    | 1.1      |
| SD     | 4   | 13  | 14  | 1.2 | 95  | 38   | 4    | 114        | 110    | 46  | 6    | 0.04  | 0.11    | 0.1     | 0.13     |
| Minimum| 3.1 | 15  | 20  | 0   | 86  | 10   | 3.7  | 137        | 113    | -130| -21  | 0.07  | 0.64    | 0.03    | 0.83     |
| Maximum| 18  | 72  | 75  | 6   | 518 | 136  | 19   | 620        | 628    | 50  | 0.3  | 0.25  | 1.1     | 0.4     | 1.3      |

Mean and median NOₓ(i), Res., and ratios to NOₓ are taken from individual data points (e.g. not mean of means).
Fig. 9. Graphs of aggregated data for the NOy(i) versus NOy where the following represent data for in balance (cross), out of balance (circle), and cloud-pumped plumes (square) (see text); and where (a) NOx versus NOy (in balance slope = 0.077 ± 0.003 with $r^2 = 0.99$ and out of balance slope = 0.028 ± 0.004 with $r^2 = 0.93$); (b) PAN versus NOy (in balance slope = 0.44 ± 0.09 with $r^2 = 0.80$ and out of balance slope 0.49 ± 0.04 with $r^2 = 0.97$); with (c) HNO3 versus NOy (in balance slope = 0.25 ± 0.03 with $r^2 = 0.92$ and out of balance slope = 0.051 ± 0.027 with $r^2 = 0.43$). Horizontal and vertical bars represent plus and minus one standard deviation about the mean for each aggregate.

$\Sigma$NOy(i) (see Figure 13c). This might suggest that the more reactive forms of NOy (i.e., NOx and PAN) may have been in sink with the average degree of atmospheric processing that had occurred in the middle/lower troposphere over these regions, whereas the substantial budget deficit for NOy - $\Sigma$NOy(i) may represent compounds that were significantly out of sink with the relative degree of atmospheric processing that is indicated by the ratios of C2H3/CO.

Ridley [1991] presented arguments that the abundance of a missing compounds implied from comparison of NOy - $\Sigma$NOy(i) are not enhanced during air mass aging. Our results agree with this argument. However, our results also indicate that the relative abundance of the implied missing compounds increases with the degree of atmospheric processing (see Figure 13d). This might be expected if the implied missing compounds are less susceptible than NOx, PAN, and HNO3 to atmospheric loss processes (e.g., oxidative attack in the case of NOx, thermal decomposition in the case of PAN, and the final dry/wet deposition processes that remove the HNO3 formed from NOx and its reservoirs).

The in-balance mixing ratios of NOy are also correlated on average with the degree of atmospheric processing established by the ratio C3H8/C2H6 (see Figure 14a). However, this trend tends to disappear in the out-of-balance data (cf. Figure 14a and 14b). This supports the argument that the various relative "chemical clocks" of the different air masses represented by these data were all perhaps on somewhat different cycles that reflect different degrees of mixing and photochemical ageing. We believe this is borne out by the loss in correlation between the ratios C3H8/C2H6 and C2H2/CO in going from the in-balance subset of data to the out-of-balance subset of data (cf. Figure 15a and 15b).

In general, our attempts at directly investigating the possible factors that influence the residuals have been complicated by (1) the apparent random resetting of selected surrogate "chemical clock" relationships, which appears especially pronounced in the out-of-balance data and (2) the generally random nature of the in-balance residual values that appears to be dominated by (1) the random uncertainties of the measurements and (2) the random uncertainties introduced by ambient variability in conjunction with the less than unity temporal overlap between the various NOy(i) and the NOy measurements, including perhaps sporadically enhanced p-NO3 events occurring within the long sample time of the aerosol measurements.

In addition to O3 mixing ratios (see earlier discussion) the only other compounds that appear to correlate with the values of the out-of-balance residuals are the mixing ratios of PAN and perhaps NOy and HNO3 (see Figure 16a-16c). The correlation found between values of the residuals and the mixing ratios of O3 and PAN would, again, suggest that the compounds comprising the measured NOy budget deficit might have some degree of covariance with the factors influencing the oxidative potential of these air
masses. These trends might support Jacob et al.'s [1992] argument that in the summertime middle free troposphere over Alaska an additional reservoir-derived source of NOx, besides that produced from the thermal decomposition of PAN, is necessary to account for the observed mixing ratios of NOx. If the NOx budget deficits represented compounds that are capable of generating NOx, then they could also enhance the photochemical production rate of O3 and possibly reinforce the observed O3 to NOx trends. However, it is doubtful that these missing compounds are simple alkyl nitrates based on the small fraction observed thus far for these compounds relative to NOy or even PAN [Atlas et al., 1992; Buhr et al., 1990].

On average, the residual values are also negatively correlated with temperature (see Figure 16d). This is consistent with there being labile reactive odd nitrogen compounds other than PAN, such as H2NO2, that are thermally stable in the cold upper troposphere [e.g., Logan et al., 1981]. Even though thermally stable at high altitudes, H2NO2 has not been predicted to be a major NOy(i) species in the summertime high-latitude middle troposphere [Singh et al., 1992; Jacobs et al., 1992]. However, their estimates did not take into account the uncertainty associated with the equilibrium H2O + NO2 = H2NO2 which is large enough to result in an approximate fivefold uncertainty in calculated mixing ratios of H2NO2 at 260 K [DeMore et al., 1992]. At the upper limit of this range, steady state mixing ratios of H2NO2 could exceed those of NOx by a factor of 10. At this limit, H2NO2 would be a major component of the NOy budget and might explain a portion of the NOy budget deficit observed in the high-latitude middle troposphere. However, before we can attribute H2NO2 to a missing portion of the NOy budget, the formidable question as to whether or not H2NO2 survives beyond the Teflon prefilter and can be quantitatively sampled by MC/IC technique must be answered.

Fig. 10. Graphs of aggregated data for O3 versus NOy where the solid line is for all nonplume points and dashed line is for NOy < 450 pptv. (a) all data (solid line slope = 0.057 ±0.005 with r² = 0.97, dashed line slope = 0.082 ±0.003 with r² = 0.97); (b) in balance (solid line slope = 0.057 ±0.009 with r² = 0.87, dashed line slope = 0.097 ±0.007 with r² = 0.98); and (c) out of balance (solid line slope = 0.059 ±0.007 with r² = 0.90, dashed line slope = 0.051 ±0.013 with r² = 0.80). Horizontal and vertical bars as in Figure 9.

An alternative explanation for these trends that we must also address is the possibility of one or more of the NOy(i) or NOx measurements is in error. Some types of errors in one or more of these measurements would still be consistent with the out-of-balance residual values being in some way correlated with factors influencing the oxidative potential of the air masses. In the case of NOy, it is unlikely that these measurements are in error by the very large factors (i.e., tenfold) that would be necessary to explain the budget deficit. This is based on the level of agreement, which has been shown in intercomparison programs, between this technique and the others (see discussion in section 2.1). In addition, potential NOy measurement interferences have a tendency to have positive values (e.g., H2NO2 -> NOy), which would argue that the NOy values might already represent upper limits. In addition, the NOy/NOx and PAN instruments were intercalibrated using common NIST NO and NO2 standards. This should have significantly reduced the potential for differences in primary standards being the cause of the NOy budget deficit.
In the case of PAN the necessary measurement errors would need to be of the order of factors of 1.5 to 3 (see Figure 16a). During the CITE 2 intercomparison, at times, 1.8-fold differences were observed between the data from the two airborne PAN instruments [Gregory et al., 1990c]. A possible error of this magnitude in the PAN measurements would also be consistent with the behavior of the missing NO$_x$ source suggested by Jacob et al. [1992]. However, it is noted that the PAN instrument used in the ABLE 3 studies was the one that on average reported larger (not smaller) PAN mixing ratios during the CITE 2 intercomparison. This would seemingly argue against the PAN measurements presented here being too small by factors of 2 to 3. In addition, we believe that the free tropospheric values of PAN that would be implied from the out-of-balance residual = PAN are unprecedentedly large.

An underestimate of HNO$_3$ could also be consistent with the observed trends. The 1988 MLOPEX results indicated an on average correlation between increasing HNO$_3$ mixing ratios and increasing values of the missing compounds that were implied by their NO$_y$ budget analysis. The out-of-balance ABLE 3 data exhibit far less evidence of this tendency (cf. Figure 16c to Figure 6 in the work of Atlas et al. [1992]). These differences between the MLOPEX and
In principle, if a compound is merely converted to another detectable gas phase NOy compound (e.g., NO3(g) + surface \rightarrow NO2(g)), or HO:NO2(g) + surface \rightarrow NO2(g)), then the effective NOy transmission efficiency of the inlet could remain high. However, if a gas phase NOy compound reacts to form a new compound that can remain on the surface, or a compound that is more easily chemiadsorbed onto a surface (i.e., sticky compounds like HNO3, e.g., HO:NO2(g) + hydrated surface \rightarrow HNO3(surface)), then the effective transmission efficiency of the inlet would decrease until all active sites are depleted, or a new temperature- and humidity- dependent equilibrium is established between the gas phase and the surface/substrate (in the case of permeable materials). As conditions change (e.g., temperature, pressure, humidity, and surface characteristics), a new set of equilibrium concentrations would also become established, which could either reduce the effective NOy transmission of the system or artificially enhance the transmission. These effects could also apply to deposited NO3−-containing aerosols and any aerosol reservoir of large polyfunctional organic odd-nitrogen containing compounds (compounds which would
ratios in order to minimize these potential problems. At this time though, we cannot rule out the possibility that the combined magnitude of these effects may be responsible for some of the random variations apparent in the in-balance data's residual/NOy statistical distribution, or some of the variance observed in previous airborne intercomparison studies (see earlier discussion in sections 2.1-2.3 and Gregory et al. [1990 c, d]). However, as far as the clearly out-of-balance data are concerned, it is difficult for us to explain this magnitude of budget deficit on the subtle sampling problems discussed above. In addition, it is difficult to understand how such effects would retain an on average tendency for correlation with O3, PAN, and temperature.

The catalytic conversion of N2O, X-NH, and X-CN compounds is believed to be small based on laboratory test. However, exhaustive tests have not been carried out using the wide range of possible conditions and chemical mixtures encountered throughout the troposphere. If catalytic conversion of these types of compounds did occur, then another possible explanation for the apparent NOy escape detection by the p-NO3 measurement technique). This combination of memory effects might generate both positive and negative artifacts that are difficult to quantify or simulate in the laboratory. In addition, the efficiency of transmitting NO3-containing or any R-NOx-containing aerosols may vary as a function of pressure (i.e., altitude) and flow velocity depending on the sampling inlet geometry and sampling line characteristics. Similar arguments can also be made for the various NOy(i) sampling systems. As noted in sections 2.1 and 2.3, some evidence of memory effects have been observed in both the NOy and the HNO3 measurement systems and may have affected some of the data presented here, especially when PFA Teflon tubing was used. All of the effects discussed above might result in somewhat random variations in instrument response. These variations could depend on the exact nature of not only the compounds constituting NOy but also ambient conditions of temperature, pressure, and humidity and the condition of the inlet sampling line's surfaces. The HNO3 and NOy sampling inlets were, however, designed to have short sample residence times and small surfaces to volume.

Fig. 14. Graphs of aggregated data for C3H8/C2H6 versus NOy: (a) in balance (cross) (slope 0.00019 ±0.00001 with \( r^2 = 0.97 \)) and (b) out of balance (circle) (slope = 0.000045 ±0.000018 with \( r^2 = 0.51 \)). Horizontal and vertical bars as in Figure 9.

Fig. 15. Graphs of aggregated data for C3H8/C2H6 versus C2H2/CO: (a) in balance (cross) (slope 0.156 ±0.023 with \( r^2 = 0.90 \)) and (b) out of balance (circle) (slope = 0.074 ±0.033 with \( r^2 = 0.56 \)). Horizontal and vertical bars as in Figure 9.
budget deficit might exist. In particular, conversion of compounds such as X-CN or X-NH could also be consistent with the correlative trends in O₃ and PAN, if these compounds became eventually oxidized to produce NOₓ. Biomass burning, which impacted all of the ABLE-3 study regions, has been shown to produce X-CN compounds [Lobert et al., 1990]. Therefore this, speculation, that of X-CN compounds, could also be consistent with the ABLE 3A ground-based NOₓ measurement results. The concurrent ground-based NOₓ, HNO₃, and NOₓ measurements in the ABLE 3A program showed good agreement in both the magnitude and the partitioning of NOₓ with the lowest altitude (~ 0.15 km) measurements taken on the aircraft [Bakwin et al., 1992]. They attributed nearly half of the ground-based measured NOₓ to "missing" compounds (~100 pptv) and very little to the labile compound PAN. Their tower flux measurements also indicated that these missing compounds also exhibited a small deposition velocity (i.e., sixfold smaller than O₃). This would suggest that the missing NOₓ compounds in the lower-altitude range are relatively unreactive. Subtle changes in the catalytic converter Au-surface's condition might significantly reduce the conversion efficiency for these more difficult to convert compounds (i.e., X-CN) without producing a noticeable change in the conversion efficiency of reducible compounds such as HNO₃, NO₂, or PAN. This type of mechanism could possibly explain the observed shift in NOₓ budget trends for the predominantly in-balance northern Labrador/Quebec data. This hypothesis would also be consistent with the suggestions that some portion of the NOₓ budget deficit is comprised of relatively unreactive nitrogen-containing compounds. The above hypothesis suggests that more detailed conversion efficiency tests need to be carried out under a wide range of conditions for a number of compounds.

4. SUMMARY

The summertime partitioning and budget of tropospheric N₂O₅ compounds have been investigated in the middle to lower troposphere over three high-latitude regions of North America. These investigations are the first to use a spectroscopically selective measurement system to detect the NO produced from a NOₓ Au-catalytic converter. The N₂O₅ budget was analyzed for the degree of closure based on the balance obtained from comparing observed NOₓ.
mixing ratios to the sum of individual NOy(i) species (NOx, HNO3, PAN, and PPN). Within the estimated precision and accuracy of the measurements, statistically significant differences were found between the ABLE 3 study regions.

In the middle free troposphere (3 - 6 km) over Alaska and the Hudson Bay lowlands, approximately 73% of the analyzed measurements indicate a deficit in the NOy budget at the 95% confidence limit. The NOy budget within the lower altitudes (0.15 - 3 km) also had a deficit. In these cases, approximately 87% of the Alaska and 35% of the Hudson Bay lowland measurements indicate a deficit at the 95% confidence limit.

Over the northern Labrador/Quebec region the NOy budget was more completely accounted for by the measured NOy(i) species. In this case, approximately 83% of the analyzed measurements fall inside the 95% confidence interval, with a nearly equal distribution about zero of residual values defined by NOy - ΣNOy(i). A combination of measurement uncertainty and poor temporal overlap between the various NOy(i) measurements is indicated as the dominant causes limiting our ability to assess the degree of closure within the NOy budget at values of ΣNOy(i)/NOy greater than about 0.6. This result is similar to that observed in other NOy to ΣNOy(i) budget studies [e.g., Parrish et al., 1993]. Where the improved level of uncertainty that has recently been observed at two of the ground-based rural locations is perhaps not unexpected when the larger NOx mixing ratios and larger fraction of more reliably measured NOx are taken into account, along with the more slowly changing atmospheric conditions found at ground-sites.

The large deficit in the NOy budget (having ΣNOy(i)/NOy < 0.6) over regions of Alaska and the Hudson Bay lowlands appears to be related to factors that influence the oxidative potential of the troposphere over these regions as reflected by the mixing ratios of O3. Based on our discussion of possible measurement errors and the correlative tendencies exhibited with other compounds, it is unlikely that the bulk of this NOy budget deficit is due to a simple artifact of the measurements. A portion of this NOy budget deficit is also suggested to be in the form of relatively unreactive nitrogen-containing compounds, especially at the lower altitudes. However, based on the correlations with O3 and temperature, a larger fraction of the implied NOx budget's missing compounds exhibit trends that are more indicative of the labile NO2-reservoir PAN. This could suggest that these missing compounds either have chemical characteristics that are similar to PAN or that they have a similar atmospheric source as PAN.

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