Field evaluation of a low-powered, profiling $pCO_2$ system in coastal Washington

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Abstract

Summertime upwelling of deep, corrosive waters on the continental shelf of the northern California Current System can exacerbate ocean acidification conditions, providing unsuitable environments for development of calcifying organisms and finfish that are important to the local economy. To better understand the carbonate system in this dynamic region, two recently developed technologies were combined with other sensors to obtain high-frequency carbon profile data from July 2017 to September 2017. The compact, low-power sensor package was composed of an optical sensor for partial pressure of carbon dioxide ($pCO_2$ optode, Aanderaa model #4797) integrated onto a wave-powered PRofi ling crAWLER (PRAWLER). The PRAWLER profiled from 3 to 80 m, stopping at fixed depths for varying lengths of time to allow for $pCO_2$ equilibration. $pCO_2$ derived from a regional empirical algorithm was used to correct optode drift using data at 80 m. Near-surface adjusted optode $pCO_2$ agreed within 6 to 42 μatm to surface $pCO_2$ from a nearby Moored Autonomous $pCO_2$ instrument. Throughout the water column, optode $pCO_2$ compared to algorithm $pCO_2$ within $-28$ to $66$ μatm. Overall, optode uncertainty was $35$ to $72$ μatm based on root-mean-square errors from all comparison data sets. Errors are attributed to optode calibration, adjustment, algorithm uncertainty, and environmental variability between optode and reference data. Improvements for optode performance within this profiling application include using more stable sensing foils, in situ calibration, and pumped flow over the sensing foil. Additionally, the study revealed undersaturated (corrosive) waters with respect to aragonite below 60 m throughout the deployment that reached up to 40 m by mid-September.

Over the last 250 yr, anthropogenic carbon dioxide (CO2) absorbed by the surface ocean has led to a measurable decrease in pH, or increase in acidity, in a process known as ocean acidification (OA) (Caldeira and Wickett 2003; Feely et al. 2004; Orr et al. 2005; Doney et al. 2009). During this time, surface pH in the open ocean has decreased by ~ 0.11 pH units and is expected to decrease by another 0.3–0.4 units before the end of the century (Orr et al. 2005). OA causes a decrease in carbonate ion concentration and calcium carbonate saturation states, which directly affects the ability of calcifying marine organisms to build their shells and skeletons (Feely et al. 2004).

On local scales and particularly in the coastal ocean, it is challenging to quantify OA and its effects on marine ecosystems due to large variability of natural and anthropogenic physical and biogeochemical processes (Bauer et al. 2013). Coastal regions that experience upwelling may be particularly vulnerable to intensified OA because of the especially high CO2 in upwelled waters compared to background values (Feely et al. 2008, 2016a, 2018; Fassbender et al. 2011; Harris et al. 2013). Because of high rates of respiration, which increases CO2 and decreases pH and dissolved oxygen ($O_2$), the deep, upwelled waters expose organisms to multiple stresses from both OA and hypoxia (Chan et al. 2008; Reum et al. 2015; Siedlecki et al. 2016). Additionally, nutrient runoff into coastal waters can exacerbate stress from natural variability, causing enhanced eutrophication and even stronger effects from hypoxia and OA (Cai et al. 2011; Wallace et al. 2014).

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Despite its importance to the economy and ecology that society values, the coastal ocean remains largely undersampled considering its high variability through space and time. In order to study dynamics of the carbonate system in the coastal ocean, in situ sensors are needed to provide accurate, high-resolution measurements of inorganic carbon parameters and pH. Traditional shipboard samples only provide snapshots of the state of the carbonate system. With autonomous, in situ platforms and sensors, spatially and temporally resolved carbon data will enable us to better assess and understand OA in the coastal ocean, at the convergence of processes such as eutrophication, upwelling, and respiration.

Over the past decade, there have been many improvements to technologies that measure the seawater carbonate system to help quantify OA. The seawater carbonate system can be fully characterized by measuring or estimating two of the four following parameters: pH, partial pressure of CO$_2$ (pCO$_2$), total alkalinity (TA), and dissolved inorganic carbon (DIC). pCO$_2$ and pH autonomous sensor technologies are relatively mature (DeGrandpre et al. 1995; Seidel et al. 2008; Martz et al. 2010, 2015; Sutton et al. 2014); however, they are often restricted to surface platforms with an infrastructure that can accommodate the sensors’ typically large physical size and significant power and reagent requirements. Thus, there is a need for more compact and low-power full-depth technologies that are still capable of maintaining climate-quality standards of measurement uncertainty (Newton et al. 2015, http://www.goa-on.org/resources/plans_strategies.php).

In this study, we combine two recently developed, low-powered technologies to observe near-full water column CO$_2$ and aragonite saturation state dynamics over the summer upwelling season in the coastal waters along the U.S. West Coast near La Push, Washington. Specifically, we deployed and evaluated the performance of a prototype pCO$_2$ optical sensor, or optode, on a wave-powered “profiling crawler,” or PRAWLER, in a novel application by parking the PRAWLER at specific depths on a mooring line to allow for equilibration of the optode sensor. Results from this study have the potential to not only help us better understand and monitor corrosive upwelling events, but also to be integrated into regional models to predict future corrosive events.

**Materials and procedures**

**Study site**

The Washington continental shelf, located in the northern part of the California Current System, contains the Olympic Coast National Marine Sanctuary (Fig. 1a). This region is a highly productive, nitrate-limited system, and even more productive than the Oregon and California shelves to the south (Hickey and Banas 2008; Alford et al. 2012). Compared to lengthy time series that exist in Oregon and California (Bograd et al. 2008; Chan et al. 2008; Nam et al. 2011), the Washington shelf is relatively understudied, in part due to extreme winter conditions that make sustained observations challenging. Steady northwesterly winds typically occurring from April to September cause warmer, fresher, low-nutrient surface waters to be transported offshore, driving local upwelling of colder, saltier, high-nutrient waters from depth. These nutrients help support large phytoplankton blooms, which then fuel the productive ecosystem. The cumulative effect of respiration in these originally deeper waters results in the upwelled high-nutrient waters also having lower O$_2$ concentrations, lower pH, and higher DIC concentrations. Upwelling, therefore, can contribute to enhanced OA and hypoxic events. The upwelled waters are then subject to additional respiration.

![Fig. 1.](https://example.com/fig1.png)

(a) Location of study site (red square) off the coast of La Push, WA. Black line is the border of the Olympic Coast National Marine Sanctuary. (b) Inset of study site, which includes the Chābā surface mooring, the Northwest Enhanced Moored Observatory (NEMO) subsurface profiler, and PRAWLER, which was deployed from 11 July 2017 to 23 September 2017 at 47.9633°N, 124.9633°W in 106 m of water.
on the shelf that subsequently reduces the O$_2$ concentrations to hypoxic levels during periods throughout the summer and early fall months (Connolly et al. 2010). Monitoring of acidified and hypoxic events has been especially important to the regional shellfish industry over the last decade after severe die-offs of larval oysters in Pacific Northwest hatcheries (Barton et al. 2012, 2015) and predicted declines in Dungeness crabs (Hodgson et al. 2018), which are economically significant fisheries to the region.

**Moorings and sensors**

The location was chosen for the optode test bed due to its proximity to the Northwest Enhanced Moored Observatory (NEMO), which is part of the Northwest Association of Networked Ocean Observing Systems (NANOOS, www.nanoos.org), and includes both surface and subsurface components (Alford et al. 2012, Fig. 1a). During this study, there were three fixed assets deployed at this location within 0.4–1.3 km of each other: the Cházhba surface mooring, the NEMO subsurface profiler mooring, and the PRAWLER mooring (Fig. 1b). The Cházhba surface mooring measures both atmospheric and oceanographic properties. Relevant to this study, within the upper meter of the surface, there is a conductivity-temperature-depth (CTD) package (SBE 16 plusV2, Sea-Bird Scientific) measuring salinity, temperature, and O$_2$, as well as a Moored Autonomous pCO$_2$ system (MAPCO2™, Battelle Memorial Institute Seaology® pCO$_2$ monitoring system) and a Submersible Autonomous Moored Instrument for pH (SAMI-pH, Sunburst Sensors) (Seidel et al. 2008; Sutton et al. 2014, 2016). Approximately 1 km to the northeast on the NEMO subsurface mooring, there is a McLane Moored Profiler (McLane Research Laboratories), which moves up and down between 20 and 90 m depth along the mooring line and measures salinity, temperature, and O$_2$ as well as other oceanographic properties not relevant to this study.

NOAA Pacific Marine Environmental Laboratory (PMEL) developed the PRAWLER (Osse et al. 2015) and transferred the technology to McLane Research Laboratories, making it commercially available in 2019. For this study, the PRAWLER was deployed on 11 July 2017 and recovered on 16 October 2017. The PRAWLER is a small, 15 kg profiling instrument that uses surface wave action to ratchet up the mooring line from depth. It then free-falls at 30 cm s$^{-1}$ to record a profile, making it capable of obtaining 20–30 profiles a day with 2–3 m vertical resolution. The PRAWLER includes real-time, two-way inductive and Iridium communication for control and data transmission to shore. It provides greater vertical sampling resolution than multiple sets of instrumentation fixed at specific depths along the mooring line and provides antifouling protection by staying out of the euphotic zone for the majority of the deployment. The tradeoff is that fixed-depth sensors can achieve higher temporal resolution at a specific depth. The PRAWLER also reduces the power load and cost of moored platforms. In addition to free-fall profiling mode, a fixed-depth mode was developed for this deployment to allow for equilibration of the pCO$_2$ optode (Model 4797, Aanderaa Data Instruments). The PRAWLER also contains a modified CTD (Sea-Bird Scientific, repackaged SBE37), as well as a dissolved O$_2$ optode (Model 4330, Aanderaa Data Instruments). There was an additional instrument package below the deepest PRAWLER depth at 88 m containing a CTD (SBE 16 plusV2) with O$_2$ (Model 4831, Aanderaa Data Instruments) and pCO$_2$ optodes, as well as a SAMI-pH (Sunburst Sensors) (Fig. 2a,b). In addition to the fixed-depth assets, a Wave Glider equipped with an Autonomous Surface Vehicle CO$_2$ (ASVCO2™) system, effectively a repackaged MAPCO2™ (C. L. Sabine unpubl.), was driven near the location to provide additional validation measurements on 12 August 2017.

Detailed description of the sensing and operational principles of the pCO$_2$ optode can be found in Atamanchuk et al. (2014). Briefly, dissolved CO$_2$ diffuses from surrounding water through a gas-permeable membrane into the sensing layer of the optode, where pH is modified relative to the pCO$_2$ level outside the membrane. Change in fluorescent properties of the pH indicator inside the sensing layer is detected by the optode as a change in phase shift ($\varphi$) of sine wave-modulated light. The optode is capable of measuring a range of 0–50,000 µatm. Atamanchuk et al. (2014) reported response times ($t_{63}$) varying from 45 s at 40°C to 4.5 min at 0°C in a stirred beaker. Reported precision was ± 2–3 µatm and absolute accuracy was 2–75 µatm (based on field data). Stability was shown to be longer than 7 months (Atamanchuk et al. 2014).

The pCO$_2$ optode was preconditioned by soaking for 2 weeks at a salinity of 34.2–34.4 and then calibrated at 21 points (seven pCO$_2$ concentrations and three temperatures) targeting the range of 250–1200 µatm at 5–25°C using a temperature-controlled water tank that was saturated with gas of varying CO$_2$ concentrations. Saturation values were checked against an independent G2201-i Analyzer CRDS CO$_2$ analyzer (Picarro). A 3D calibration model is calculated where log(pCO$_2$) is fit to a sixth by second degree polynomial that is dependent on $\varphi$ and temperature. The initial calibration for the optode was later expanded to cover a broader range of pCO$_2$ values, because pCO$_2$ > 1200 µatm (the initial upper limit) was observed during the deployment. This procedure included fitting a second-degree polynomial to initial $\varphi$ and pCO$_2$ data at each temperature ($R^2 = 0.9999$), adding calibration points for $\varphi$ to increase the calculated pCO$_2$ range up to 2000 µatm, and updating calibration coefficients accordingly. The expansion of the initial calibration polynomial could be a source of error that will be described in a later section. After the optode is deployed, the sensing foil requires in situ conditioning with ambient seawater, which is a process that occurs on the order of days (Atamanchuk et al. 2014, 2015a,b).

**Validation measurements**

For surface validation, pCO$_2$ measured by the optode on the PRAWLER (henceforth referred to as PRAWLER optode pCO$_2$; Table 1) was compared to the measurements from the
MAPCO2™ system on the Čážba mooring that sampled at ~0.5 m depth every 3 h (henceforth referred to as Čážba-MAPCO2™; Table 1). The PRAWLER surface depth was between 3 and 6 m (due to range in swell height which caused the PRAWLER to heave up and down while clamped to the mooring line). In addition, PRAWLER optode pCO2 measurements were compared to surface pCO2 measured at ~0.5 m depth and within <2 km by a Wave Glider ASVCO2™ system (henceforth referred to as ASVCO2™; Table 1).

For subsurface validation, we used empirical relationships developed for the northern California Current System, which are based on similar methodology to Alin et al. (2012), to estimate carbonate system parameters from 42°N to 50°N, 0 to 200 km offshore, and 0 to 100 m depth. A key assumption in the empirical algorithms is that the stoichiometry of the relationship between O2 and DIC is primarily controlled by aerobic respiration in the water mass since it was last exposed to the atmosphere. Additionally, it is assumed that upwelling is the main process that brings these water masses to the continental shelf. Here the algorithm was extended to the surface; however, the different exchange rates of CO2 and O2 at the air-sea interface may result in a larger error in estimates generated in the upper 15–30 m of the water column by using the relationship between CO2 and O2. Surface heating or cooling

Fig. 2. (a) Schematic diagram of PRAWLER buoy and instrumentation located above the acoustic release. (b) Instruments deployed on the PRAWLER mooring: (1) PRAWLER with CTD, O2 optode, and pCO2 optode (white arrow points to optode position on PRAWLER, the casing was removed prior to deployment); (2) Seabird 16 CTD with O2 and pCO2 optodes; and (3) SAMI pH and external battery.
Table 1. Data sets assessed for pCO₂ optode performance and their description. See Fig. 1 for platform locations.

| Data set name | Data set description |
|---------------|----------------------|
| PRAWLER pCO₂ | pCO₂ measured by the Aanderaa pCO₂ optode on the PRAWLER (either raw measurements or adjusted using algorithm reference data) |
| Čážba-MApCO₂™ | pCO₂ measured by the MApCO₂™ system on the Čážba- surface buoy |
| ASVCO₂™ | pCO₂ measured by the ASVCO₂™ system on the surface wave glider |
| Modeled PRAWLER pCO₂ | pCO₂ calculated by algorithm-derived DIC (Eq. 1) and TA (Eq. 2) using T, S, and O₂ measurements from the PRAWLER |

will also change how temperature affects the relationships. The empirical algorithms (Eqs. 1, 2) use oxygen (O₂ in μmol kg⁻¹), temperature (T in °C), salinity (S), and the potential density anomaly (σθ in kg m⁻³) to calculate DIC and TA.

\[
\text{DIC } \mu\text{molkg}^{-1} = 955.6636926 - 0.776753102*\text{O}_2 + 51.19911384*\sigma_\theta \tag{1}
\]

\[
\text{TA } \mu\text{molkg}^{-1} = 673.4949547 + 44.53461935*S + 0.496611464*T + 278.56666762*1/T \tag{2}
\]

These relationships achieved R² values of 0.97 and 0.90 and root-mean-squared errors (RMSEs) of 18 μmol kg⁻¹ and 12 μmol kg⁻¹ for DIC and TA, respectively, for the calibration data set of National Oceanic and Atmospheric Administration Ocean Acidification Program West Coast Ocean Acidification (NOAA OAP WCOA) cruises from 2007 to 2016 (Feely and Sabine 2013; Feely et al. 2015, 2016b,c; Alin et al. 2017). Estimated DIC and TA were then used to calculate pCO₂ concentrations using CO2SYS (Lewis and Wallace 1998; van Heuven et al. 2011; Orr et al. 2018) using carbonate (Lueker et al. 2000), sulfate, and borate equilibrium constants (Dickson 1990a,b). Phosphate and silicate equilibrium constants are from Millero (1995) and total boron is calculated using Uppstrom (1974). pCO₂ calculated from the algorithms using PRAWLER hydrographic data will be referred to as modeled PRAWLER pCO₂ (Table 1). Modeled PRAWLER pCO₂ had an average estimated uncertainty of ~16% based on propagating the RMSE for TA and DIC in the CO2SYS calculation of pCO₂ (Orr et al. 2018). The average error of the modeled PRAWLER pCO₂ relative to high-quality Čážba-MApCO₂™ (< 2 μatm uncertainty) was found to be 13%, which is within the model uncertainty. Note that only data from the most equilibrated PRAWLER periods when the PRAWLER and Čážba- buoys were measuring the same source waters (defined here as ΔT < 0.5 and ΔS < 0.5) was used.

Discrete bottle samples were taken the day after the mooring was deployed. However, the pCO₂ optode was not fully conditioned such that the bottle samples could not be used as a reference for the optode measurements. Prior to recovery of the instrument package, the PRAWLER ran out of power before it was possible to take validation bottle samples. Therefore, DIC and TA from collected bottle samples were used to compare to algorithm-calculated DIC and TA using measured parameters during sample collection. The resulting RMSE between the modeled and bottle data for DIC was larger than the model suggests at 26 μmol kg⁻¹ instead of 18 μmol kg⁻¹, but lower than the model TA RMSE at 5 μmol kg⁻¹ instead of 12 μmol kg⁻¹ (n = 16 for DIC and TA, from sampled depths 2–87 m). Although we intended to also compare the pCO₂ and pH measured by the instrument package at the bottom of the mooring line, there were multiple instrument failures due to a faulty voltage channel on the CTD, and the data were unrecoverable.

Assessment

PRAWLER vehicle performance

The carbon PRAWLER package was deployed on 11 July 2017 and recovered on 16 October 2017. The battery on the PRAWLER lasted until 23 September after conducting 614 profiles over 74 d. The PRAWLER obtained 7–8 profiles per day with 0.5–1 h between profiles and each profile took 2–3 h to complete. As mentioned previously, the PRAWLER was used primarily in fixed-depth mode during this deployment. There were nine depth stations set at approximately 10 m intervals (4, 10, 20, 30, 40, 50, 60, 70, and 80 m), although actual depth of sampling could vary by 1–4 m depending on swell height. At each depth station, the PRAWLER would first be parked at the specified depth and not record data, a so-called soak period, and then sample continuously at an 8 s interval for 8 min. After finishing the sampling period, the PRAWLER would move down to the next depth. After reaching the bottom, the PRAWLER would ratchet up and down the mooring line until it was time to begin the next profiling cycle starting at the surface. With two-way iridium communication, we had the ability to receive data from the PRAWLER and change any of the sampling settings in real time from shore.

Observed patterns in PRAWLER T, S, and O₂ measurements

Data from the first 2 weeks of the deployment were not considered in the data analysis due to optode in situ conditioning, which is typically on the order of days (Atamanchuk et al. 2014, 2015a,b). Over the remaining deployment period, temperature ranged between 7°C and 17°C with temperatures >10°C found in the top 20 m (Fig. 3a,b). Salinity in near-surface (3–6 m) waters was as low as 30.5 in the beginning of the deployment and increased to ~32 by the end of July, and stayed within 32 ± 0.5 until the end of the deployment (Fig. 3c,d). Salinity increased with depth and was ~34 at the 80 m depth station throughout the deployment. There was strong stratification due
to the warmer, fresher waters on the surface and colder, saltier waters at the deeper depths.

In addition to temperature and salinity effects on oxygen solubility and mixing with upwelling waters, O₂ concentrations were also influenced by biological processes such as production and respiration. Dissolved O₂ concentrations ranged approximately from 300 to 500 μmol kg⁻¹ in near-surface (3–6 m) waters (Fig. 3e,f). This translates to 80–200% O₂ saturation, where O₂ was supersaturated (> 100%) for 75% of the deployment. At ~ 80 m, O₂ concentrations were between 80 and 120 μmol kg⁻¹ until 10 September, and afterward, values decreased to 5–100 μmol kg⁻¹ translating to 1–35% saturation. These O₂-depleted waters (< 120 μmol kg⁻¹) reached depths as shallow as 40 m toward the end of the deployment, where hypoxic waters (< 60 μmol kg⁻¹, shown in dark red Fig. 3e,f) are present through the rest of the deployment. Cold, low-oxygen water intrusions are apparent in approximately week-long periods in the deeper water column throughout the deployment. Periods of hypoxic waters were also seen at 85 m in September 2014 and 2016 at the nearby Chāba mooring (PSEMP Marine Waters Workgroup 2017). These low O₂ waters were often associated with southerly winds and currents and advected northward along the shelf based on data collected on the ASVCO2™ as well as the NEMO and Olympic Coast National Marine Sanctuary buoys (PSEMP Marine Waters Workgroup 2017). These observations were seen previously in this region by Connolly et al. (2010) in 2005 and by the Puget Sound Marine Waters Workgroup in last several years.

**Optimal equilibration time for pCO₂ optode**

Response time of the pCO₂ optode is mainly affected by temperature, salinity, and flow rate over the sensing foil. Temperature is a primary factor affecting response time of the optode: diffusion of CO₂ into the sensor foil is slower in colder water. The response time (τ63) reported in Atamanchuk et al. (2014) was 4 min at 0.5°C, meaning that 99% equilibration would take ~ 18 min, while response time at 20°C was
The temperature is measured with a built-in sensor on the optode as well as with the CTD on the PRAWLER. The optode may also drift out of calibration temporarily after extended exposure to salinity that is > 1 salinity unit different from the salinity used for optode calibration due to the so-called “osmotic effect,” which is a result of the osmotic pressure between different salinities. This issue was avoided by parking the optode at depth and close to its calibration salinity of 34.2–34.4 when not profiling. While each profile took 2–3 h, there was not a sustained period of time in a significantly different salinity and, therefore, it did not affect the optode calibration. Pressure effects are negligible at <100 m. Storage and conditioning affect accuracy of optode measurements, but do not influence response time, that is, speed of equilibration.

The first objective for the deployment was to find the optimal amount of time needed to reach full equilibration for the pCO2 optode. In this study, we tested soak times of 0, 2, 4, and 10 min followed by sampling at an 8 s interval for 8 min. Data presented herein (exclusive of Fig. 4) are the mean of the last minute of sampling at each parked depth, which contained the last seven measurements. This method used the most equilibrated data available. Total time was calculated as soak time in addition to sampling time, where the longest equilibrated measurement at a specific depth was 18 min. From 02 to 09 September, there was no soak time, or in other words, the PRAWLER was in free-fall/profiling mode (see full profile data in Fig. 3a,c,e), which did not provide usable data for the pCO2 optode (this will be discussed further in the next section). Free-fall profiling mode does, however, provide reliable T, S, and O2 data (Fig. 3a,c,e).

The pCO2 optode data were categorized by periods of soak time (2, 4, or 10 min) to find the appropriate length of time needed at a fixed depth to achieve full equilibration. For each soak time in surface (0–5 m), middle (15–45 m), and deep (55–85 m) depths, pCO2 was averaged over the 8-min sampling period and the average pCO2 difference (ΔpCO2) was found by subtracting the first average pCO2 from the average pCO2 at each 8-s sampling interval (Fig. 4). Negative ΔpCO2 values are seen in the surface due to the large vertical pCO2 gradient where the optode was exposed to higher pCO2 in the colder, deeper waters before equilibrating to much lower pCO2 in the warmer surface waters (Fig. 4a). Throughout the water column, the 10-min soak time data had the smallest range in average ΔpCO2 and therefore, shallowest slopes, while the 2-min soak time data had the largest average ΔpCO2 and steeper slopes (Fig. 4). This is because the disequilibrium for pCO2 is higher after the 2-min soak time than after the 10-min soak time. For example, for the deep bin, where average temperature was ~8°C, fitting an exponential to each equilibration curve led to 72%, 76%, and 89% equilibration after 2-min, 4-min, and 10-min soak times, respectively. It would have taken 37–40 min for 99% equilibration, which is almost double the time of ~21 min for 99% equilibration in a stirred beaker at 0.5°C (Atamanchuk et al. 2014), and 55–59 min for 99.9% equilibration.

Faster equilibration would be expected for warmer temperatures in surface waters; however, during this deployment, there was faster equilibration at deeper, colder depths (Fig. 3a, b). This could be because the optode equilibration time is affected by a combination of temperature and pCO2 concentration gradients. Both absolute temperature values as well as
temperature gradients are important to consider in their effects on optode response time for this deployment. At depths shallower than 20 m, there is a larger temperature gradient, while at depths greater than 20 m, temperature was approximately the same. In between measurement profiles, the PRAWLER was moving up and down the mooring line to maintain the optode calibration by not sitting outside its calibration salinity for a sustained period. This may have led to insufficient time, especially at shorter soak times, for the optode to equilibrate at the surface due to the significant temperature change between the surface and subsurface depths. Due to the lack of sufficient active flow over the membrane and the membrane proximity to the titanium sensor body, the pCO₂ optode may have developed a memory effect of colder, higher pCO₂ bottom waters.

Each soak period reached at least 63% equilibration at middle and deep depths; however, if we define “fully” equilibrated as within 2 μatm for the last minute, the only fully equilibrated data at middle depths is the 10-min soak time and at deeper depths, the 4- and 10-min soaks. Hence, a 10-min soak time, or 18-min equilibration gave the most equilibrated data, at middle depths is the 10-min soak time and at all points in the 10-min sections not including after 13 September (Fig. 5e, Table 2), For all of the data at 80 m including the 2- and 4-min soak periods, the mean residual was $-5 \pm 39$ μatm or $-0.6\%$ with $n=1454$ and $R^2=0.98$, suggesting that the assumption of a linear drift between 10-min sections was valid. These adjustments were then expanded to all depths to provide the adjusted PRAWLER optode pCO₂ that will be reported from here on.

Although Chába- pCO₂ surface data could have also been used as reference data to adjust the PRAWLER optode pCO₂, we did not choose this method due to multiple factors: (1) If Chába- pCO₂ were used as reference data, the mean residual between Chába- pCO₂ and PRAWLER optode pCO₂ (adjusted to Chába-) would be $-2 \pm 32$ μatm ($n=200$) (not shown), which is similar to $-1 \pm 35$ μatm ($n=508$) using the 80 m algorithm but the algorithm provides a larger number of comparison data points (Fig. 5); (2) The algorithm performs adequately in the surface where modeled Chába- pCO₂ (the algorithm is calculated with input parameters from the Chába- mooring) compared to measured Chába- pCO₂ resulted in average offset of 22 ± 18 μatm or an 8% error (Fig. 5c), which is within the stated uncertainty of the model at 16%. This offset and scatter is likely due to biological activity and different air-sea gas exchange rates in the surface ocean impacting reliability of the algorithm. In addition, both the model and Chába-suggested similar drift in the uncorrected or raw optode measurements at 80 m and the surface, respectively ($\gamma = -3.5x + 104.9$ and $\gamma = -5.6x - 44.5$, where $x$ is the number of days since deployment) (Fig. 5a,d); (3) There is a drift of $-2 \pm 32$ μatm d⁻¹ between the Chába- and adjusted PRAWLER pCO₂ suggesting that there is still environmental variability between the Chába- and PRAWLER moorings that is not accounted for by considering when locations have similar temperature and salinity. Using the 80 m algorithm-based estimates as a reference eliminates spatial differences and environmental variability between the Chába- and PRAWLER mooring locations (0.4 km away from each other) and differences in sampling depths (0.5 m for Chába- and 3–6 m for PRAWLER).

**Optode adjustment using algorithm reference data**

Following procedures described by Atamanchuk et al. (2014), we used modeled PRAWLER pCO₂ from 80 m as reference data to calculate an offset adjustment to the calibration model that uses log(pCO₂) as a function of phase shift and temperature. This adjustment corrects for any conditioning, storage-related, or instrument drift. At 80 m, the algorithm model is not degraded by surface ocean biological productivity and air-sea gas exchange. Hence, raw optode (phase shift $\phi$) measurements were taken from the uncorrected PRAWLER optode pCO₂ at 80 m, and an adjustment was made to minimize the average residual between PRAWLER optode pCO₂ and modeled PRAWLER pCO₂ at the same depth and time (Fig. 5d,e). Only sections of the most equilibrated (10-min soak) data were used for this adjustment. A correction was applied to each of the three sections of 10-min soak data excluding the low O₂/high CO₂ period after 13 September (Period 1: 03–17 August, Period 2: 28 August–02 September, and Period 3: 08–13 September, see Figs. 5, 6), and a linear drift over time was assumed for the adjustments for data between 10-min sections. Data were interpolated to hourly increments for comparison purposes. Before adjustment, the mean residual between the uncorrected PRAWLER optode pCO₂ and modeled PRAWLER pCO₂ was $-265 \pm 83$ μatm or $-18\%$, suggesting a fairly large offset and scatter as well as an overall drift of ~5.5 μatm a day from the first day of deployment (Fig. 5d). After adjustment for each 10-min soak period was corrected for the drift and offset, the mean residual was $-1 \pm 35$ μatm or $-0.2\%$ with $n=508$ and $R^2=0.96$ for all points in the 10-min sections not including after 13 September (Fig. 5e, Table 2). For all of the data at 80 m including the 2- and 4-min soak periods, the mean residual was $-5 \pm 39$ μatm or $-0.6\%$ with $n=1454$ and $R^2=0.98$, suggesting that the assumption of a linear drift between 10-min sections was valid. These adjustments were then expanded to all depths to provide the adjusted PRAWLER optode pCO₂ that will be reported from here on.

**Optode pCO₂ observations and comparison to other observations/models**

After allowing for 2 weeks of conditioning for the optode from 11 July to 24 July (data not shown), adjusted PRAWLER optode pCO₂ ranged approximately from 70 to 760 μatm in near-surface waters (Figs. 6c, 7a). Values under 200 μatm fall
below the calibration lower limit of the optode and therefore may be less accurate. At 80 m, concentrations ranged from 1000 to 2000 $\mu$atm. The highest $p_{CO_2}$ values at 80 m occurred in the second to last week of the deployment with sustained values > 1500 $\mu$atm. This occurs simultaneously with the arrival of low O$_2$ waters (Fig. 3) due to the advection of hypoxic waters from the south.

The $p_{CO_2}$ observations from the Chába MAPCO2™ and ASVCO2™ have the lowest uncertainty of the validation data sets (Sutton et al. 2014), and therefore, were the primary

**Fig. 5.** $p_{CO_2}$ measured during the three sections of 10-min soak data excluding the low O$_2$/high CO$_2$ period after 13 September (Period 1: 03–17 August, Period 2: 28 August–02 September, and Period 3: 08–13 September). Chába MAPCO2™ data are either measured and modeled using algorithms and only from periods when the two locations are measuring similar water masses ($\Delta T < 0.5$ and $\Delta S < 0.5$). PRAWLER optode data are shown as raw measurements, modeled using algorithms, and adjusted per period using 80 m reference algorithm data. Surface comparisons are shown in (a–c) and 80 m comparisons are shown in (d, e). Residuals (black) are shown with a line of best fit. The x-axis is the number of days since the deployment (11 July 2017) to show how the measurements drift after the deployment.
focus for estimating the total error in the adjusted PRAWLER optode pCO₂ data. However, significant variability in the PRAWLER data hinders comparison among data sets. Surface seawater temperature (SST), salinity, and pCO₂ data collected from the PRAWLER are much noisier than Chába data (Fig. 6). One likely factor is the differences in the sampling locations, including differences in surface depth of measurements (3–6 m for the PRAWLER and 0.5–1 m for Chába) and the distances between the two moorings (0.4 km; Fig. 1b). The high near-surface stratification and heave of the PRAWLER, in combination with the slow equilibration time of the optode, all play a role in the larger variability of the PRAWLER data. The impact of equilibration time on the pCO₂ optode is also apparent during the section of the deployment with a 2-min soak time. The shorter equilibration time clearly causes increased variability in pCO₂ from the optode (Fig. 6c,d).

Despite the noise in the data, similar patterns are seen in both data sets. Sea surface temperature at Chába is at the high end of the PRAWLER SST, which corresponds to Chába measuring shallower, warmer surface water. Sea surface salinity (SSS) is the opposite, where Chába SSS is at the low end of the PRAWLER SSS, likely because the Chába CTD was measuring shallower surface water that was also less saline. During the period from 13 to 18 September, the PRAWLER optode pCO₂ increases approximately from 400 μatm from the week before to a peak of ~ 800 μatm, which is a similar magnitude of increase as the Chába MAPCO₂ data ~ 300 μatm to a peak of ~ 600 μatm, or approximately a doubling of absolute surface seawater pCO₂ values. During this period, surface PRAWLER optode pCO₂ measurements are routinely higher than Chába, which could be partially a result of measuring

Fig. 6. Surface data plotted from the PRAWLER and Chába moorings as well as the Wave Glider ASVCO₂™ (a) SST, (b) SSS, (c) surface seawater pCO₂, and (d) pCO₂ residuals between the PRAWLER (adjusted to 80 m algorithm data) and Chába. The shaded region indicates the PRAWLER free-falling profile period, for which the pCO₂ data are not shown due to insufficient equilibration. The 2, 4, and 10 min labels in (c) refer to the soak times for each period of time.
slightly deeper, cooler, and more saline water; however, the optode is measuring the same proportion of change. This large pCO₂ increase could be a result of the upwelling of CO₂-rich, hypoxic waters advected from the south.

In order to estimate PRAWLER optode pCO₂ error and uncertainty, we control for both environmental variability between the PRAWLER and Chaiba buoys as well as the period of short equilibration time. First, as a control to equilibration time, we only consider Chaiba- and PRAWLER data when the SST and SSS suggest both locations are sampling the same water mass (ΔSST < 0.5 and ΔSSS < 0.5). In addition, we only consider data from 10-min soak periods and exclude the high CO₂ and low O₂ period after 13 September. Data are interpolated to hourly intervals to enable better comparison among the three measurements of the PRAWLER, Chaiba, and ASVCO₂. The average residual between Chaiba- MAPCO₂™ and adjusted PRAWLER optode pCO₂ (or pCO₂ from Chaiba – pCO₂ from PRAWLER optode adjusted to the 80 m modeled reference data) is 6 ± 42 μatm or 2% error (n = 200; Table 2). ASVCO₂™ pCO₂ data taken from within <2 km of the PRAWLER mooring have an average residual of 35 ± 31 μatm or 16% error (n = 10) compared to PRAWLER optode pCO₂.

Although the modeled pCO₂ uncertainty is much larger than the primary validation data sets from the MAPCO₂™ and ASVCO₂™, the full profile data from the model provide another independent approach for evaluating the PRAWLER optode pCO₂. Modeled PRAWLER pCO₂ (Fig. 7c,d) follows similar patterns as the PRAWLER optode pCO₂ (Fig. 6a,b) and ranged approximately from 130 to 560 μatm in the surface waters to ~2000 μatm at depth. Values above the threshold for hypercapnia (pCO₂ > 1000 μatm, outlined in black Fig. 7b,d; McNeil and Sasse 2016; Feely et al. 2018) are present in the bottom water through most of the deployment. The highest concentrations of pCO₂ > 1500 μatm are seen at depth toward the end of the deployment and co-occur with a low O₂ (Fig. 3) and high surface pCO₂ signal (Fig. 6). Taking a closer look at residuals between the optode and model pCO₂ (Fig. 7e,f), there are a few key areas where the optode and model differ significantly. Apart from the free-falling profiling period and the last 2 weeks of the deployment, the PRAWLER optode pCO₂ shows values lower than modeled PRAWLER pCO₂ in the top 20 m with positive residuals. The differences in exchange rates of O₂ and CO₂ at the air-sea interface that would alter the linear relationship between O₂ and CO₂ assumed in the empirical algorithm model (Alin et al. 2012) can explain the observed differences between the model and optode pCO₂. Since the optode is adjusted using data at 80 m, the relationship between O₂ and CO₂ in deep water that is largely influenced by respiration may not be as accurate for near-surface waters where processes such as photosynthesis and air-sea gas exchange drive concentrations of O₂ and CO₂.

Overall, depths below 20 m show a neutral to slightly negative residual, indicating the optode pCO₂ measurements tend to be higher than modeled pCO₂. This is particularly apparent in the last 2 weeks of the deployment (during the period of time that experienced advection of hypoxic waters from the south), which shows stronger negative residuals through most of the water column. This result is not surprising given that low-oxygen water masses advected from the south tend to have undergone significant in situ respiration on the shelf after their initial upwelling onto the shelf (Connolly et al. 2010; Siedlecki et al. 2015, 2016). Thus, the fundamental relationships underpinning the algorithm have been altered under these conditions.

For the 10-min soak periods (excluding the last 2 weeks of Period 3 that contained the low O₂ waters), the average residual between PRAWLER optode pCO₂ and modeled PRAWLER pCO₂ was −28 ± 66 μatm (n = 1337), or −1%, and RMSE of 72 μatm. This RMSE is larger than errors between the pCO₂

| Depth          | Distance to PRAWLER mooring (km) | Validation data set | Average residual ± SD compared to PRAWLER optode pCO₂ (μatm) | % Difference | RMSE (μatm) | n  |
|----------------|----------------------------------|---------------------|--------------------------------------------------------------|-------------|-------------|---|
| Surface (0.5 m) | 0.4                              | Chaiba-MAPCO₂™      | 6 ± 42                                                       | 2           | 43          | 200 |
| Surface (0.5 m) | <2                               | ASVCO₂™             | 35 ± 31                                                      | 16          | 46          | 10  |
| Water column (3–80 m) | 0                              | Modeled PRAWLER pCO₂ | −28 ± 66                                                     | −6          | 72          | 1337 |
| Bottom (80 m)  | 0                                | Modeled PRAWLER pCO₂ | −1 ± 35                                                      | −0.2        | 35          | 508 |
optode and MAPCO2™ and ASVCO2™ observed at the surface (Table 2), which is likely due to additional algorithm errors. Additionally, this larger error could be affected by the expansion of the optode calibration polynomial to accommodate $pCO_2$ values $>1200$ μatm, values that are seen primarily at depths $>60$ m. Therefore, the uncertainty of 72 μatm based on the RMSE is likely an upper bound estimate because it includes adjustment and algorithm errors as well as environmental factors.

Applying PRAWLER measurements to calculate aragonite saturation state

Aragonite is a calcium carbonate mineral that many organisms use to form their shells. We can use aragonite saturation state ($\Omega_{ar}$) as an indicator of ecosystem health and vulnerability to OA, particularly for shell-forming organisms (Bednaršek et al. 2017). The ability to measure aragonite saturation state variability in the Northern California Current, both spatially and temporally, is of great interest to local fisheries of oysters, Dungeness crabs, et al that have strong sensitivities to $\Omega_{ar}$, especially when $\Omega_{ar}$ reaches below species-specific biological thresholds. Here we compare calculations of $\Omega_{ar}$ using two methodologies in order to determine feasibility of different sensor packages for observing OA on a PRAWLER. First, $\Omega_{ar}$ is calculated using an empirical algorithm similar to those for DIC (Eq. 1) and TA (Eq. 2) based on methodology by Alin et al. (2012), using temperature and dissolved O2 concentrations (Eq. 3, Fig. 7a):

$$\Omega_{ar} = -17.167825 + 1.52647616 \times T + 0.00547706 \times O_2 + 60.5017384 \times 1/T - 0.038633 \times T^2 $$

This multiple linear regression results in an $R^2$ value of 0.95 with RMSE value of 0.15 when compared to the calibration data set of the NOAA OAP WCOA cruises from 2007 to 2016 (Feely and Sabine 2013; Feely et al. 2015, 2016b,c; Alin et al. 2017). Second, $\Omega_{ar}$ is also calculated using $pCO_2$ measured...
from the optode and the empirical algorithm for TA from Eq. 2. Uncertainty is calculated for $\Omega_{st}$ by propagating the respective uncertainties for $pCO_2$ and TA through CO2SYS calculations applied to the whole data set (Orr et al. 2018). Using an average optode uncertainty from the three validation sets in Table 2 of 51 $\mu$atm, estimated $\Omega_{st}$ uncertainty is 0.36 using Eq. 2 TA. Thus, the algorithm-based $\Omega_{st}$ achieves lower uncertainty than the $\Omega_{st}$ calculated using PRAWLER optode $pCO_2$ and estimated TA, therefore the rest of this section will be presented using algorithm-based $\Omega_{st}$ results.

Using the algorithm for direct calculation, $\Omega_{st}$ ranged from close to 0 to ~ 4 (Fig. 8a,b). $\Omega_{st}$ values > 1.5 were only seen in the surface waters shallower than 20 m depth. Below 60 m, undersaturation ($\Omega_{st} < 1$) was observed throughout the deployment with periods where undersaturation extended to as shallow as 40 m. Extremely low $\Omega_{st} < 0.5$ is found at 80 m depth in mid-September, coinciding with low $O_2$, high $CO_2$ seen previously (Figs. 3, 7). The NOAA OAP WCOA cruises (Feely et al. 2015, 2016b,c; Alin et al. 2017) sampled at various times between June and September of 2011 to 2013 and 2016. DIC and TA discrete samples were used in CO2SYS to calculate $\Omega_{st}$ from the closest sampling stations from WCOA cruises (< 27 km away) resulting in undersaturated waters only seen below 80 m depth. From June 2016 at < 1 km away from the PRAWLER location, aragonite undersaturation started at ~ 83 m.

**Discussion**

**Field evaluation of $pCO_2$ optode performance**

To briefly summarize the optode performance in the field compared to our “gold standard” methods of Chârba-MAPCO2™ and ASVCO2™ for surface data, the adjusted PRAWLER optode $pCO_2$ data have an average residual of 6 ± 42 $\mu$atm and 35 ± 31 $\mu$atm with RMSEs of 43 $\mu$atm and 46 $\mu$atm as well as 2% and 16% difference, respectively (Table 2). Using algorithm-estimated data to assess error throughout the water column, we see a slightly larger error and RMSE for optode data compared to the model of $-28 \pm 66 \mu$atm, $-6\%$ difference, and RMSE of 72 $\mu$atm. Besides sensor error, spatial heterogeneity of water properties across sites contributes to the difference between PRAWLER optode and Chârba-MAPCO2™ and ASVCO2™, while model uncertainty and on-shelf processes that can alter the fundamental relationships underlying the DIC model contribute to estimated error of the PRAWLER $CO_2$. In conclusion, the total uncertainty estimate based on the RMSEs for the $pCO_2$ optode during the 2-month deployment on the PRAWLER of 43–72 $\mu$atm is likely conservative as it includes both sensor error and environmental variability between the PRAWLER and reference measurements. Still, our results are comparable to the reported performance of the $pCO_2$ optodes in previous studies, where observed accuracy varied within 2–75 $\mu$atm of the true value during deployments of 6–7 months in the dynamic coastal environment (Atamanchuk et al. 2015a,b). In these studies, optode measurements were compared to calculated $pCO_2$ from DIC and TA water samples, with errors attributed to incomplete sensor conditioning, sensor drift, calculation, and water sampling. In another study, a $pCO_2$ optode was deployed at a fixed depth side-by-side with an oxygen probe in Lake Illmensee, Germany for 2 months in order to investigate lake metabolism (Peeters et al. 2016). The $pCO_2$ data from the optode were compared against $pCO_2$ data from a reference $CO_2$-IR probe (HydroC™ CO2, CONTROS) and were corrected for conditioning drift using the same methodology applied in...
this study. Good agreement of the amplitude and the timing of the daily fluctuations in $pCO_2$ measured with the CO$_2$ optode and the CO$_2$-IR probe was found after the adjustment. On average, $pCO_2$ data had an error of $-1 \pm 12$ $\mu$atm compared to CO$_2$-IR during a $\sim 5$ d period when both probes were deployed side-by-side. After application of the scaling factor to compensate for the initial conditioning drift, the $pCO_2$ optode provided reliable data over the duration of the study (Peeters et al. 2016). Better performance of the optode in the study of Peeters et al. (2016) can be attributed to the lack of dynamic errors introduced by profiling mode in this study and to the smaller range of $pCO_2$ values, where Peeters et al. (2016) covered a $pCO_2$ range of $\sim 100$ $\mu$atm and this study covers a range of $\sim 2000$ $\mu$atm. Additionally, the study by Peeters et al. (2016) was conducted in a lake with freshwater, so there were no salinity changes.

**Comments and recommendations**

The PRAWLER proved to be an effective test bed vehicle to use for our goal of evaluating prototype Aanderaa $pCO_2$ optode performance off the Washington coast. Its ability to transmit data in real time back to shore and allow for modifications to the sampling scheme were instrumental to this type of sensor assessment. Additionally, temperature and oxygen measured on the PRAWLER enabled us to estimate aragonite saturation with half the uncertainty than using the optode. In order to measure $pCO_2$ to weather- or climate-quality, target precision and accuracy would be 2.5% or 0.5%, respectively. A target response time would be similar or faster than available oxygen optodes, which ranges from 6 to 25 s (Bittig et al. 2014), in order to deploy the same sampling schemes and achieve the same resolution for the two parameters. The optode achieved an error uncertainty of 43–72 $\mu$atm based on RMSE (Table 2) during the 2-month deployment on the PRAWLER. It did not lose its calibration throughout the study. Better performance of the optode in the study of Peeters et al. (2016) can be attributed to the lack of dynamic errors introduced by profiling mode in this study and to the smaller range of $pCO_2$ values, where Peeters et al. (2016) covered a $pCO_2$ range of $\sim 100$ $\mu$atm and this study covers a range of $\sim 2000$ $\mu$atm. Additionally, the study by Peeters et al. (2016) was conducted in a lake with freshwater, so there were no salinity changes.

Suggestions to improve optode performance on future deployments involve more stable and sensitive sensing foils, reliable in situ calibration, and adding a pump to provide faster seawater flow rates against the sensor foil for faster response times. Sensor drift should be monitored by comparing against reference data such as discrete bottle samples or the multiple checkpoints per day that we had planned to have using the instrument package below the mooring line. Future long-term profiling deployments would benefit from a more accurate instrument colocated with the optode that does not need to achieve the same measurement frequency (Peeters et al. 2016). Faster flow rates could be achieved by adding a pumped flow of water or a funnel could be attached to direct water to flow over the sensor foil which would increase the flow rate without requiring additional power. To allow the optode temperature to equilibrate faster and reduce the thermal mass, different housing materials could be used for the optode body. Additionally, for future deployments on the PRAWLER, surface waters may require longer soak times to allow the optode to equilibrate temperature and $pCO_2$, but shorter times could be used for deeper waters where temperature is more uniform. Faster equilibration and improved accuracy in $pCO_2$ optode measurements would provide high-frequency data that could be used to calculate calcium carbonate saturation states to provide further insights into dynamics of OA. These results could be incorporated into regional and global biogeochemical models to improve predictions of corrosive events.

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Acknowledgments
The authors would like to thank the captain and crew of R/V Jack Robertson and F/V Aquila for helping with the deployment and recovery of the PRAWLER mooring. We would also like to thank John Shanley, Keith Magness, Ryan Newell, Zoe Parsons, Chris Archer, Morgan Langis, and others in the NOAA PMEL Carbon and Engineering groups for their efforts in collecting these data. This work was funded by the NOAA Ocean Acidification Program and the Joint Institute for the Study of the Atmosphere and Ocean (JISAO) under NOAA Cooperative Agreement NA15OAR4320063. This is JISAO Contribution No. 2018-0182 and PMEL Contribution No. 4907.

Conflict of Interest
None declared.