Drastic decline of flood pulse in the Cambodian floodplains (Mekong River and Tonle Sap system)

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Received: 17 July 2021 – Discussion started: 21 July 2021
Revised: 19 December 2021 – Accepted: 4 January 2022 – Published: 7 February 2022

Abstract. The Cambodian floodplains experience a yearly flood pulse that is essential to sustain fisheries and the agricultural calendar. Sixty years of data, from 1960–2019, are used to track the changes to the flood pulse there. We find that minimum water levels over 2010–2019 increased by up to 1.55 m at Kratie and maximum water levels decreased by up to 0.79 m at Prek Kdam when compared to 1960–1991 levels, causing a reduction of the annual flood extent. Concurrently, the duration of the flooding season has decreased by about 26 d (Kampong Cham) and 40 d (Chaktomuk), with the season starting later and ending much earlier. Along the Tonle Sap River, the average annual reverse flow from the Mekong to the Tonle Sap Lake has decreased by 56.5 %, from 48.7 km³ in 1962–1972 to 31.7 km³ in 2010–2018. As a result, wet-season water levels at Tonle Sap Lake dropped by 1.05 m in 2010–2019 compared to 1996–2009, corresponding to a 20.6 % shrinkage of the lake area. We found that upstream contributors such as current hydropower dams cannot fully account for the observed decline in flood pulse. Instead, local anthropogenic causes such as irrigation and channel incision are important drivers. We estimate that water withdrawal in the Cambodian floodplains is occurring at a rate of (2.1 ± 0.3) km³ yr⁻¹. Sediment decline and ongoing sand-mining operations have also caused channel erosion. As the flood pulse is essential for the ecological habitats, fisheries and livelihoods of the region, its reduction will have major implications throughout the basin, from the Tonle Sap system to the Vietnamese Mekong Delta downstream.

1 Introduction

The Mekong River in Southeast Asia has attracted much attention as water infrastructure development has accelerated in the past years (Best, 2019; Soukhaphon et al., 2021). Due to its transboundary nature, the cross-border hydrological impacts of anthropogenic alterations have become a contentious topic (Stone, 2010). For instance, Keovilignavong et al. (2021) described how the recent 2019–2020 Mekong drought was politicised by states and agencies.

The Cambodian floodplains and the Tonle Sap Lake system is home to a unique geographical phenomenon. During the dry season, the lake empties into the Mekong. However, during the wet season, large tracts of the floodplains are inundated, and flow is reversed from the Mekong to the lake. This annual flood pattern is critical for both the productivity of fisheries (Halls and Hortle, 2021; Sabo et al., 2017; Ziv et al., 2012) and the agrarian communities that are reliant upon the annual floodwaters for replenishment of nutrients and water (Arias et al., 2012; Grundy-Warr and Lin, 2020).

Due to the significance of the annual floods, the ecological and hydrological services provided by the Cambodian floodplains is best understood as a consequence of the flood pulse (Junk et al., 1989). Elsewhere, hydrological alterations to flood pulses have been quantified in the Amazon Basin (Zulkafi et al., 2016) and Missouri Basin (Bovee and Scott, 2002). Within other parts of the Mekong, the flood pulse has been investigated vis-à-vis its relationships to climate (Räsänen and Kummu, 2013; Västilä et al., 2010) and ecosystems.
(Arias et al., 2013; Kong et al., 2017; Ngor et al., 2018). These studies indicate that quantification of flood pulses can be useful in the understanding of the hydrology of a floodplain system.

In the Cambodian floodplains, the water levels in the Tonle Sap Lake and the Mekong mainstream have a close relationship (Guan and Zheng, 2021; Inomata and Fukami, 2008), meaning that any alterations to flows on the mainstream will affect the hydrology of the lake. Studies of the Tonle Sap Lake have predicted that dry-season water levels will increase and wet-season water levels will decrease (Arias et al., 2012; Kuumu and Sarkkula, 2008). Indeed, remotely sensed data has confirmed that the surface area of the lake has shown a declining trend since 2000 (Ji et al., 2018; Lin and Qi, 2017).

Overall, the Cambodian floodplains can protect neighbouring Phnom Penh and the Vietnamese Mekong Delta (VMD) downstream from floods by storing large volumes of water during the wet season (Fujii et al., 2003; Mekong River Commission et al., 2004). In addition, the effects of upstream water infrastructure development, such as the Lancang dams, on the VMD are also dampened by the buffering effect of the floodplains (Dang et al., 2015). However, despite its importance, the hydrological changes experienced by the Cambodian floodplains over the past decades have been poorly understood.

This study first aims to quantify the flood pulse along the Cambodian floodplain using sixty years of data, from 1960–2019. Then, we want to identify the anthropogenic factors that have caused the alterations to the flood pulse. Through this approach, we offer novelty in two ways. First, we studied the Cambodian floodplains in its entirety, as compared to other authors who only investigated the Tonle Sap system (Chen et al., 2021; Kuumu and Sarkkula, 2008) or the Mekong system (Binh et al., 2020b). Second, we synthesised knowledge of the various anthropogenic drivers in the Cambodian floodplains and associated them with observed hydrogeomorphological impacts. In so doing, we present the implications of current human activities on the Cambodian floodplains and the wider region.

2 Study area

Beginning from Stung Treng, the lower reaches of the Mekong winds through a large floodplain that is seasonally inundated during the wet season (Fig. 1b). Connected to the floodplain is the Tonle Sap system, which expands from an area of 2500 km$^2$ during the dry season to up to 15 000 km$^2$ during the wet season (Ji et al., 2018). The whole floodplain is underlain by a mosaic of tropical ecosystems such as gallery forest, shrublands and aquatic herbaceous vegetation (Araki et al., 2007; Arias et al., 2013; Kuumu and Sarkkula, 2008), and man-made land use such as rice fields and canals (Mahood et al., 2020; Olson and Morton, 2018). When water level exceeds 12 m at Kampong Cham, extensive overbank flooding will occur on both banks of the Mekong (Inomata and Fukami, 2008; Mekong River Commission et al., 2004). Concurrently, along the Tonle Sap River, water flows from the Mekong to the Tonle Sap Lake, bringing along its supply of sediment and nutrients (Campbell et al., 2009; Lu et al., 2014b; Siev et al., 2018). In this study, unless otherwise defined, “upstream” refers to stretches of the Mekong located above Stung Treng.

We identify three main anthropogenic factors that can cause changes to the yearly flood pulse: upstream dam construction, irrigation and sand mining. In the Upper Mekong Basin (UMB), the Lancang cascade, consisting of 11 dams over an 800 m drop, was built by China beginning with Manwan Dam in 1992 (Hecht et al., 2019). These dams have raised concerns due to their ability to alter the hydrological regime downstream (Lu et al., 2014a). For example, the two largest, Xiaowan and Nuozhadu, completed in 2010 and 2014 respectively, have a total reservoir capacity of 38.3 km$^3$, which is more than half of the total capacities of all reservoirs across the whole Mekong Basin (Mekong Region Futures Institute, 2020). The cascade has been found to increase dry-season discharge and reduce wet-season discharge downstream (Li et al., 2017; Räisänen et al., 2012) to as far as Kratie (Räisänen et al., 2017). Nearer the Cambodian floodplains, tributary dams such as the Pak Mun Dam in the Chi-Mun system have also been shown to regulate downstream Mekong flows (Cochrane et al., 2014). Dams in the 3S (Sekong, Sesan, Srepon) river basin can also affect the flood pulse downstream by increasing minimum water levels (Arias et al., 2014) or by decreasing wet-season discharge (Piman et al., 2013a).

In recent years, the Cambodian floodplains have been developed to tap into their potential for rice production (Erban and Gorelick, 2016; Yu and Fan, 2011). Under the “Rice-White Gold” policy of the Cambodian government, the country fulfilled its objective to increase rice yield to 4 million tonnes in 2015 from just 20 000 tonnes in 2009 (Royal Government of Cambodia, 2010). The boom in rice production was possible with both the use of high-yielding rice varieties and an expansion of irrigation infrastructure that allowed dry-season cropping (Asian Development Bank, 2019b). With investments from international donors such as the World Bank and the Asian Development Bank (ADB) and countries such as Japan and China, Cambodia has both upgraded its ageing irrigation schemes and constructed new canals and reservoirs (Sithirith, 2017; Table S1 in Supplement). However, this water infrastructure consumes large volumes of water – Erban and Gorelick (2016) estimate that full dry-season irrigation could use up to 31% of the total Mekong and Bassac flow.

Another anthropogenic activity that has increased pace is sand mining within the river channels (Kondolf et al., 2018; Schmitt et al., 2017). In Cambodia alone, it was estimated that at least 34.4 million cubic metres of sediment
Figure 1. (a) Map of the Mekong Basin showing sites of major mainstream water infrastructure as of 2019. (b) Map of the Cambodian floodplains, beginning from Stung Treng to the Vietnamese border. The floodplains also include the Tonle Sap Lake. (c) Map of the Chaktomuk confluence between the Mekong mainstream, Tonle Sap River and Bassac River. Possible irrigation canals are also shown. The canals in grey are extracted from OpenStreetMap. © OpenStreetMap contributors 2022. Distributed under the Open Data Commons Open Database License (ODbL) v1.0.
were mined per year (Bravard et al., 2013). Of this amount, 18.1 million cubic metres was mined within the short stretch of the Mekong from Kampong Cham to the Vietnamese border (Bravard et al., 2013). As large volumes of sediment are being removed from the river bed, sediment contributions from the Upper Mekong are unable to replenish these losses (Hackney et al., 2020). The problem is so severe in the neighbouring VMD that sand mining has resulted in an up to 1.3 m decrease in water levels, exacerbating the impacts of sea-level rise and saltwater intrusion (Brunier et al., 2014; Vu et al., 2018). On the Cambodian side, Hackney et al. (2021) estimate that sand mining has caused riverbed incision at a rate of −0.26 m yr$^{-1}$ from 2013 to 2019 with no signs of abatement. Consequently, Ng and Park (2021) conjectured that this incision is causing the Tonle Sap Lake to shrink.

3 Materials and methods

3.1 Data

Water level and discharge data were obtained from the Mekong River Commission (Mekong River Commission, 2021). On the Mekong mainstream, we obtained data from Stung Treng, Kratie, Kampong Cham, Chroy Changvar and Neak Luong. We also obtained data from Prek Kdam (Tonle Sap River), Chaktomuk (Bassac River) and Kampong Luong (within Tonle Sap Lake). These stations were selected due to their good coverage of the Cambodian floodplains and documentation of their historical records from 1960 onwards (Table 1).

However, there are some gaps in the digital data downloaded from the MRC data portal. For these gaps, we consulted physical data records at the MRC office and Ministry of Water Resources and Meteorology, Cambodia. While most of these missing entries were found and filled, there are still some years without documentation despite our best efforts (Fig. S1 in the Supplement). Precipitation data were also obtained from the MRC for Kampong Cham and Chaktomuk.

At Kampong Cham, Neak Luong and Chaktomuk, missing discharge data from 2012–2019 were calibrated from their corresponding water levels using the rating curves derived by the Mekong River Commission et al. (2004; Eqs. 1 to 3). To test the validity of these curves, they were used to generate predictions of daily discharge values from 2003–2011. The predictions were then compared with actual discharge readings measured during the same period to obtain their respective $R^2$ score. The plots of actual and predicted values are available in Fig. S2 in Supplement.

Kampong Cham (KC): $Q_{KC} = (8.869H_{KC} + 29.811)^2$

$$H_{KC} - H_{CC} = 0.3 + 2412.182(R^2 = 0.998)$$ (1)

Neak Luong (NL): $Q_{NL} = (12.718H_{NL} + 62.250)^2$

$$H_{CC} - H_{NL} = 0.2(R^2 = 0.993)$$ (2)

Chaktomuk (CK): $Q_{CK} = (13.943H_{CK} + 19.992)^{1.8}

$$R^2 > 0.999$$

Here, $Q$ is discharge and $H$ is water levels. $H_{CC}$ refers to water level at Chroy Changvar.

The high $R^2$ scores indicate high accuracy of the predicted discharge values and thus justify their use for further analysis. Additionally, we identified an unreported change in the rating curve at Stung Treng. The new curve was implemented on 1 January 2005 and depressed discharge values afterwards, making them unsuitable for comparison with prior data (Lu and Chua, 2021). Therefore, post-2004 discharge data must be calibrated with the previous rating curve. The previous rating curve was obtained by plotting discharge and water level data from 2000–2004 and then seeking the best fit line with the smallest RMS error. The resulting polynomial relation (Eq. 4) achieved high accuracy with $R^2 > 0.999$:

Stung Treng (ST): $Q_{ST} = 207.549H_{ST}^2 + 2598.316H_{ST}$

$$- 4854.477(R^2 > 0.999).$$ (4)

3.2 Methodology

Figure 2 offers a timeline of major anthropogenic processes occurring in the Mekong Basin from 1960 to 2020. There are three distinct phases of water infrastructure development: 1960–1991, 1992–2009 and 2010–2019. 1960–1991 constitutes the pre-dam era and can be treated as the historical baseline before any major water infrastructure development. Following the construction of the first dam on the Mekong mainstream – the Manwan Dam – the growth era, from 1992 to 2009, saw extensive hydropower development both on the Mekong mainstream and its tributaries upstream in China, Thailand and Laos. For example, during this period, China constructed the Manwan, Dachaoshan and Jinghong dams in the Upper Mekong Basin, with a total storage capacity of no less than 2.95 km$^3$ (Mekong Region Futures Institute, 2020). Within Cambodia, irrigation infrastructure was still small scale and sand mining was only on the scale of 13.5 Mt yr$^{-1}$ (United Nations, 2017). However, after 2010, the pace of dam construction increased with the operationalisation of mega-dams in China such as the Xiaowan and Nuozhadu dams, with a combined storage capacity of up to 38.3 km$^3$, thereby marking the start of the mega-dam era (MRFL, 2020). Concurrently, the Cambodian government announced the “Rice-White Gold” policy paper in 2010, sparking a burst of intensive irrigation projects in the Cambodian floodplain (Asian Development Bank, 2019a; Royal Government of Cambodia, 2010). By 2019, the pace of sand mining increased to almost 50.2 Mt yr$^{-1}$ (Hackney et al., 2021), almost 4 times its 2009 rate.

This division of our study period into the three eras allows us to view any hydrological changes within the context of wider basin developments in the Mekong. Furthermore, this division is consistent with other studies in the region (Binh
Table 1. Summary of obtained data records from the various stations.

| Station            | Year Range  | Actual Data Water Level | Actual Data Discharge | Computed Discharge Data |
|--------------------|-------------|--------------------------|-----------------------|-------------------------|
| Mekong mainstream  | 1960–2019   | 1960–2019                | 2005–2019             | Calibrated to 2000–2004 rating curve because new rating curve depressed discharge values from 2005 onwards |
| Stung Treng       |             |                          |                       |                         |
| Kratie             | 1960–2019   | 1960–1969                | 1980–2019             |                         |
| Kampong Cham       | 1960–2019   | 1960–2011                | 2012–2019             | Used the rating curve from the Mekong River Commission et al. (2004) to estimate discharge for missing years |
| Chroy Changvar     | 2000–2019   | 1960–2011                | 2012–2019             | Used the rating curve from the Mekong River Commission et al. (2004) to estimate discharge for missing years |
| Neak Luong         | 1960–2019   | 1960–2011                | 2012–2019             | Used the rating curve from the Mekong River Commission et al. (2004) to estimate discharge for missing years |
| Tonle Sap River    | 1960–2019   | 1962–1972                | 1995–2018             |                         |
| Prek Kdam          |             |                          |                       |                         |
| Bassac River       | 1960–2019   | 1960–2011                | 2012–2019             | Used the rating curve from the Mekong River Commission et al. (2004) to estimate discharge for missing years |
| Chaktomuk          | 1960–2019   | 1960–2011                | 2012–2019             |                         |
| Tonle Sap Lake     | 1996–2019   |                          |                       |                         |
| Kampong Luang      | 1996–2019   |                          |                       |                         |

Figure 2. Timeline of major water infrastructure activities in the Cambodian floodplains and the wider Mekong Basin. Blue labels indicate mainstream dams constructed across the Mekong (including the Lancang Cascade). Black labels indicate irrigation projects within Cambodia. The red chart represent the estimated rate sand mined per year in millions of tonnes (Mt). Mining data from 1998–2015 were obtained from the United Nations (2017), while mining data from 2016–2020 were obtained from Hackney et al. (2021).
et al., 2020b; Guan and Zheng, 2021; Li et al., 2017; Räsänen et al., 2017), allowing for the cross-comparison of results. However, due to missing data years at some stations, the study duration might not be consistent. Nevertheless, our analysis still reflects the general trend and central tendency across the different timeframes because only average values were compared.

3.2.1 Flood pulse

The various parameters used to characterise flood pulse as seen in Fig. 3a were determined from Eqs. (5) to (14) below.

\[
\text{MIN}_{\text{annual}} = \text{lowest water level of a year} \tag{5}
\]

\[
\text{MAX}_{\text{annual}} = \text{highest water level of a year} \tag{6}
\]

\[
\text{AMPLITUDE} = \text{MAX}_{\text{annual}} - \text{MIN}_{\text{annual}} \tag{7}
\]

\[
\text{MIN}_{\text{annual}} (m) = 50\text{th percentile of all water levels in study period} \tag{8}
\]

\[
\text{START DATE} = \text{Date when water level} > \text{FT AND remain more than FT for next 10 d} \tag{9}
\]

\[
\text{END DATE} = \text{Date when water level} < \text{FT AND remain less than FT for next 10 d} \tag{10}
\]

\[
\text{DURATION (d)} = \text{Days between START DATE and END DATE} \tag{11}
\]

\[
\text{PEAK DATE} = \text{Date of MAX}_{\text{annual}} \tag{12}
\]

\[
\text{RISE RATE} (m d^{-1}) = \frac{\text{MAX}_{\text{annual}} - \text{FT}}{\text{Days from START DATE to PEAK DATE}} \tag{13}
\]

\[
\text{FALL RATE} (m d^{-1}) = \frac{\text{MAX}_{\text{annual}} - \text{FT}}{\text{Days from PEAK DATE to END DATE}} \tag{14}
\]

Comparisons of the means of numerical variables such as AMPLITUDE, RISE RATE and FALL RATE were conducted with Welch’s t test. For comparison of dates, the date was first converted to its Julian date. Thereafter, the median date of each study period was compared and validated with a Mann–Whitney test.

3.2.2 Reverse flow at Prek Kdam

The parameters used to characterise reverse flow (RF) where water flows from the Mekong to Tonle Sap Lake are stated in Eqs. (15) and (16):

\[
\text{ANNUALRF} (\text{km}^3) = \sum \text{RF in a year} \tag{15}
\]

\[
\text{MAX}_{\text{RF}} (\text{cms}) = \text{highest RF of a year} \tag{16}
\]

Three phases of the flow pattern are observed: normal phase, transition phase and reverse phase (Fig. 3b). The normal phase is when water flows from the Tonle Sap towards the Mekong. The transition phase begins on the date when the river record water flowing in the reverse direction for the first time in the year. The reversal phase starts when water continues to flow towards the lake and does not switch direction until the water abruptly changes back to the normal phase flow. Note: not all years have a transition phase.

3.2.3 Discharge changes

At each station, the mean discharge (in cubic metres per second) from June to September was computed. By comparing average values during the pre-dam and mega-dam eras, the change in flood pulse across the entire Cambodian floodplain can be visualised. Additionally, the difference in discharge \(Q_{\text{diff}}\) between the downstream stations of Chaktomuk (CK) and Neak Luong (NL), and the upper station of Stung Treng (ST), was calculated to estimate the amount of water loss/gain in the Cambodian floodplains (Eq. 17).

\[
Q_{\text{diff}} = (Q_{\text{CK}} + Q_{\text{NL}}) - Q_{\text{ST}} \tag{17}
\]
4 Results

4.1 Mekong River mainstream

4.1.1 Changes to annual flood extent

The annual flood extent is given by the yearly maximum and minimum water levels of the Mekong. Compared to the pre-dam era from 1962–1991, minimum water level was higher during the growth era from 1992–2009 (Fig. 4). The trend continued into the mega-dam era (2010–2019) with significant increases recorded at all stations except Chaktomuk (Table S2 in the Supplement). For reference, the minimum water levels were higher by 0.60 m at Stung Treng, 1.55 m at Kratie, 0.60 m at Kampong Cham and 0.10 m at Neak Luong.

Furthermore, maximum water levels decreased at all stations downstream of Kratie during the mega-dam era as compared to the pre-dam era. For example, Neak Luong and Chaktomuk experienced 0.55 and 0.76 m drops respectively. Correspondingly, the amplitude of the flood pulse decreased. Comparing records from the mega-dam era to the pre-dam era, the amplitude dropped by 7.9 % at Kratie, 5.6 % at Kampong Cham, 10.6 % at Neak Luong and 8.9 % at Chaktomuk. This observed increase in dry-season minima and decrease in wet-season maxima are consistent with studies in other parts of the Mekong Basin (Binh et al., 2020b; Li et al., 2017; Räsänen et al., 2017), demonstrating that the impacts of water infrastructure development are evident within the Cambodian floodplains.

In practical terms, more areas of the riverbanks are now permanently inundated during the dry season. As the flood amplitude decreases, the annual flood extent is reduced, meaning that some parts of the floodplains are no longer being flooded during the wet season.

4.1.2 Changes to annual flood duration

When compared against 1960–1991 records, the flood seasons in 1992–2009 did not show many significant changes. However, the mega-dam era showed greater alterations in flood timing. Downstream from Kratie, flood duration decreased significantly by 26 d at Kampong Cham, 36 d at Neak Luong and 40 d at Chaktomuk. The drastic shortening of the flood season by up to more than a month was caused by both a delay in flood pulse and an early end to the flooding. The delay in the start of the flooding season was observed to increase with further distance downstream. For instance, while the start date was only later by 9 d at Kampong Cham, it was later by 11 d at Chaktomuk and 15 d at Neak Luong. For the end dates of the flood season, it was earlier by 18 d at Kampong Cham, 25 d at Chaktomuk and 18 d at Neak Luong. Thus, for areas downstream of Kratie, the floodwaters have indeed arrived later and receded earlier, resulting in a shorter wet season and a longer dry season.

4.1.3 Changes to rise and fall rates

Changes in rise/fall rates reflect influences of upstream water infrastructure. During the rising limb of the wet season, reservoirs have to release the water stored during the dry season in preparation for the incoming water (Richter et al., 1997; Singer, 2007). Also, the presence of irrigation canals increases the conveyance speed of floodwaters across the floodplains, resulting in an increased rise rate. After the wet season, upstream reservoirs and irrigated fields retain water (Cochrane et al., 2014). As flows to the main channel is reduced, the fall rate is correspondingly higher.

At Kampong Cham, Chaktomuk and Neak Luong, both rise and fall rates were observed to increase during the mega-dam era as compared to the pre-dam era. The largest change in rise rate was observed at Chaktomuk with a 53.8 % increase, from 0.062 to 0.096 m d\(^{-1}\). In terms of fall rates, the largest percentage change was observed at Kampong Cham: a 23.8 % increase, from 0.089 to 0.111 m d\(^{-1}\). These observed alterations point at anthropogenic hydrological regulation in the region.

4.2 Tonle Sap system

4.2.1 Changes to water exchange at Tonle Sap River

At Prek Kdam (Fig. 4f), the hydrological changes mirrored those in the Mekong mainstream. Compared against pre-dam records, the mega-dam era saw minimum water levels increased significantly by 0.25 m and maximum water level decreased by 0.79 m. This resulted in a significant 12.9 % reduction in amplitude – from 8.27 to 7.20 m. Furthermore, the flood duration was shorter by around 20 d. The average start date in the mega-dam era was 10 July, a significant 15 d delay from 25 June previously during the pre-dam era.

The total annual outflow from the Tonle Sap Lake to the Mekong decreased from 74.54 km\(^3\) in 1962–1972 to 62.81 km\(^3\) in 2010–2019 (Fig. 5a and Table S3 in the Supplement). Similarly, the reverse flow from the Mekong to the Tonle Sap decreased from 49.67 km\(^3\) in the pre-dam era to only 31.74 km\(^3\) in the mega-dam era – a drastic reduction of 56.5 %. In addition, the duration of the reverse phase decreased by around 13 d, from 125 d in 1962–1972 to 112 d in 2010–2019.
4.2.2 Changes to Tonle Sap Lake flooding pattern

At the Kampong Luang station in the Tonle Sap Lake, annual minimum water levels decreased from 0.70 m in 1996–2009 to 0.60 m in 2010–2019 (Fig. 5b and Table S4 in the Supplement). Similarly, the yearly maximum water level dropped significantly from 8.58 m (1996–2009) to 7.52 m (2010–2019). Using the relation of water level ($H$) to volume ($V$) and area ($A$) as derived by Kummu et al. (2014; Eqs. 18 and 19),

$$A = -5.5701H^3 + 137.40H^2 + 470.29H + 1680.2(R^2>0.99)$$  \hspace{1cm} (18)

$$V = 0.7307H^2 - 0.3554H + 0.9127(R^2>0.99); \hspace{1cm} (19)$$

the 1.05 m drop in the maximum water level meant that 3990 km$^2$ of previously seasonally inundated land is now permanently dry. Correspondingly, the reduction in maximum water volume is a drastic 12.1 km$^3$. This reduction translates to a decrease in 20.6% of maximum area and 23.4% of maximum water volume. During the dry season, the minimum area is now 3.1% smaller (2080 to 2010 km$^2$) and contain 5.8% less water (1.02 to 0.96 km$^3$).

This observed decrease in both the minimum and maximum water levels is consistent with recent scholars who monitored the change in Tonle Sap Lake area using remote sensing methods (Ji et al., 2018; Lin and Qi, 2017; Wang et al., 2020). However, the observed decrease in dry-season flow does not follow hydro-modelling results that suggest that the water level should be higher during the dry season at the Tonle Sap Lake (Arias et al., 2012, 2014; Kummu and Sarkkula, 2008; Piman et al., 2013b). We postulate that the models adopted might not have considered the development...
of water infrastructure on Tonle Sap Lake tributaries (Asian Development Bank, 2019a).

Additionally, there was a significant drop in flood duration, from 198 to 163 d. This reduction was caused by an early end to the flood season. During 1996–2009, the average end date of the flood season was 31 January but during 2010–2019 the season ended roughly on 18 January, earlier by 13 d. The shortened flooding season follows the same trend as in the Mekong mainstream.

4.3 Changes to the flood pulse

Figure 6 shows a schematic of the annual flood pulse on the Cambodian floodplains. During the pre-dam era of 1960–1991, the discharge at Stung Treng during the wet season was $25,500 \text{m}^3 \text{s}^{-1}$. Downstream at Kratie, discharge reduced slightly to $24,000 \text{m}^3 \text{s}^{-1}$, which could be caused by overland flooding between Stung Treng and Kratie. From Kratie to Kampong Cham, as additional water arrived from the surrounding watersheds, discharge increased to $27,100 \text{m}^3 \text{s}^{-1}$. Thereafter, extensive overland flooding was experienced at the Chaktomuk confluence of the Mekong, Bassac and Tonle Sap rivers. At Prek Kdam, $4270 \text{m}^3 \text{s}^{-1}$ of water flowed to the Tonle Sap Lake. At Chaktomuk and Neak Luong, $2800 \text{m}^3 \text{s}^{-1}$ and $16400 \text{m}^3 \text{s}^{-1}$ of water continued to flow towards the Vietnamese Delta via the Bassac and Mekong rivers respectively.

During the mega-dam era of 2010–2019, there were reductions in wet-season discharge across all stations. At Stung Treng, discharge only decreased by $300 \text{m}^3 \text{s}^{-1}$, a small 1.2 % reduction. At Kratie and Kampong Cham, discharge decreased to $22,600$ and $21,200 \text{m}^3 \text{s}^{-1}$, a larger percentage reduction of 5.8 % and 21.7 % respectively. All subsequent downstream stations also observed lower discharge. Flow towards the Tonle Sap Lake at Prek Kdam dropped from $37.5 \%$ to $2670 \text{m}^3 \text{s}^{-1}$ and discharge at Neak Luong and Chaktomuk decreased to $14,800$ and $2200 \text{m}^3 \text{s}^{-1}$ respectively.

As seen in Fig. 7, whether $Q_{\text{diff}}$ is positive or negative depends on the time of the year. In the dry months from October to April, $Q_{\text{diff}}$ is generally positive, implying that there is a net contribution of water from the tributaries. Alternatively, in the wet months from June to September, $Q_{\text{diff}}$ is generally negative as water is lost through the increased evapotranspiration from the flooding. This alternating gain-loss pattern is elucidated in Fig. 8.

Figure 8b also showed two general trends in $Q_{\text{diff}}$ across time. During 1960–1975, $Q_{\text{diff}}$ showed an increasing trend. However, from 1976–2019, $Q_{\text{diff}}$ decreased significantly at a slope of $-(68 \pm 8) \text{m}^3 \text{s}^{-1} \text{yr}^{-1}$. This implied that the Cambodian floodplains have been losing water at a rate of $(2.1 \pm 0.3) \text{km}^3 \text{yr}^{-1}$.

5 Discussion

5.1 Impacts of upstream dams and precipitation

Section 4 has demonstrated that the flood pulse has indeed been decreasing across the Cambodian floodplains. However, whether the reduction is caused by upstream dams, climate or local operations is hotly debated. For instance, the claim that Chinese dams have minimal impact on the inundation area of the Tonle Sap by Wang et al. (2020) was refuted in a response letter by Kallio and Kummu (2021). To investigate these competing drivers, we compared the reduction of wet-season discharge across various stations.

Since Stung Treng, the uppermost station of the Cambodian floodplains, only registered a reduction of $300 \text{m}^3 \text{s}^{-1}$, then the cumulative contribution of infrastructure and precipitation in the upper reaches of the Mekong only accounted for $300 \text{m}^3 \text{s}^{-1}$ of discharge. Propagating this $300 \text{m}^3 \text{s}^{-1}$ decrease further, the percentage of discharge reduction caused by upstream development versus that by local development cannot be ascertained. Thus, only 21 % of reduction at Kratie and 5 % of reduction at Kampong Cham could be attributed to the upper reaches (Table 2). Similarly, even though the downstream stations of Prek Kdam, Neak Luong and Chaktomuk also registered declines in discharge, only about 8 % of the flow reduction at these three lower stations could be attributed to development upstream. Furthermore, there were no significant changes to either maximum flood levels or flood duration at Stung Treng (Table S2 in the Supplement).

The changes in $Q_{\text{diff}}$ further support the argument that local factors are more likely to be the main reason for the reduction in flood pulse. The only natural mechanism able to explain the reduction in $Q_{\text{diff}}$ as shown in Sect. 4.3 is that overbank flooding has become more frequent, leading to an increased loss of water through evapotranspiration. However, from the analysis of flood pulse in Sect. 4.1, we have identified that both flooding extent and duration show a decreasing trend. Therefore, the only plausible mechanism for the drop in $Q_{\text{diff}}$ is local diversion of Mekong flows.

Also, as observed in Fig. 9, measured rainfall in the Cambodian floodplains has remained roughly constant from 1960–2019, in line with observations via other sensing methods (Raghavan et al., 2018; Singh and Qin, 2020; Theune, 2015). This observation implies that local precipitation is not the driving factor for the reduction in discharge. Thus, the observed reduction of flood discharge in the Cambodian floodplains cannot be attributed solely to either upstream development or natural climatic variability – local anthropogenic factors are likely the main reason.

5.2 Impacts of water withdrawal

The scale of irrigation and construction of irrigation reservoirs has been increasing since the Cambodian government announced its “Rice-White Gold” policy with numer-
Figure 6. Schematic of wet-season discharge on the Cambodian floodplains during the pre-dam and mega-dam era. Across all stations, there is a reduction of discharge during the mega-dam era of 2010–2019.

Figure 7. Monthly plots of $Q_{\text{diff}}$ (cms) from 1960–2019. Blue zones indicate $Q_{\text{diff}} > 0$, meaning an addition of water from outside the Cambodian floodplains. Red zones indicate $Q_{\text{diff}} < 0$, meaning a loss of water to outside the Cambodian floodplains. Generally, during the wet season from June to September, $Q_{\text{diff}} < 0$. 

Hydrol. Earth Syst. Sci., 26, 609–625, 2022 🍵 https://doi.org/10.5194/hess-26-609-2022
Table 2. Percentage contribution of influences upstream of Stung Treng to the reduction of discharge at the various stations in the Cambodian floodplains.

|                      | Wet-season discharge (cms) | Total reduction (cms) | From upper reaches (cms) | Contribution (%) | From Cambodian floodplains (cms) | Contribution (%) |
|----------------------|----------------------------|-----------------------|--------------------------|-----------------|---------------------------------|-----------------|
|                      | Pre-dam                    | Mega-dam              |                          |                 |                                 |                 |
| Stung Treng          | 25 500                     | 25 200                | −300                     | −300            | 100                             |                 |
| Kratie               | 24 000                     | 22 600                | −1400                    | −300            | 21                              | −1100           | 79              |
| Kampong Cham         | 27 100                     | 21 200                | −5900                    | −300            | 5                               | −5600           | 95              |
| Downstream (Prek Kdam + Neak Luong + Chaktomuk) | 23 470                    | 19 670                | −3800                    | −300            | 8                               | −3500           | 92              |

Figure 8. (a) Time series plot of $Q_{\text{diff}}$ from 1960 to 2019 with monthly steps. (b) Time series plot of $Q_{\text{diff}}$ from 1960 to 2019 with yearly steps. From 1976–2019, $Q_{\text{diff}}$ decreased at a slope of $−(68 \pm 8) \text{ m}^3 \text{ s}^{-1} \text{ yr}^{-1}$ ($p < 0.01$).

There are two main drivers of channel incision: reduced sediment flux and sand mining. Due to hydropower development in the Upper Mekong Basin, sediment flux has declined in the past decades (Bussi et al., 2021; Kondolf et al., 2014; Kummu et al., 2010; Lu et al., 2014b; Wang et al., 2011). This sediment reduction results in the formation of “hungry water” (Kondolf, 1997), which causes greater erosion of the channels, leading to channel deepening (Kondolf et al., 2018; Lu et al., 2007). Concurrently, sand mining has increased pace within the last decade, with mining areas expanding the most rapidly within the Chaktomuk confluence area at Phnom Penh (Hackney et al., 2021). Other areas identified as mining hotspots include the stretch from Phnom Penn to the Vietnamese border along both the Mekong and Bassac channels (Bravard et al., 2013).

5.3 Impacts of channel incision

As Fig. 7 shows, there has been a sharp and obvious decline in $Q_{\text{diff}}$ during November and December from 2010 onwards, coinciding with the start of the Cambodian ‘Rice-White Gold’ policy. This reduction follows the planting calendar of dry-season cropping from around December to April (Cramb et al., 2020). While the wet-season cropping is primarily rainfed and does not require much diversion of water, dry-season cropping must be extensively irrigated (Mekong River Commission, 2009). Consequently, farmers will store water during the preceding year’s wet season for the upcoming second growing season (Phengphaengsy and Okudaira, 2008).

Although irrigation coverage has increased to approximately 22.6% of paddy land by 2015, the percentage is still much lower than in neighbouring Thailand and Vietnam (Kea et al., 2016), meaning that Cambodia can still expand its rice production. As drought occurrences are likely to increase in the future (Oeurng et al., 2019), water availability will be the major challenge for the continued expansion of the Cambodian agriculture sector (Bresney et al., 2020; Sithirith, 2021).
Figure 9. Mean 3-month rainfall at Kampong Cham and Chaktomuk from 1960–2019. There is no statistically significant trend in rainfall within the data period (Mann–Kendall test: $p = 0.8$).

Figure 10. Comparison of water levels at Neak Luong/Chaktomuk against water levels at Kratie during February (black: 1960–1991; blue: 1992–2009; red: 2010–2019). Given the same water level at Kratie, the water levels at Neak Luong/Chaktomuk have been decreasing over time.

The relationship given in Fig. 10 shows that incision at the downstream stations has occurred relatively faster than incision at upstream Kratie. If incision were faster at Kratie, then the $y$ intercept of the best-fit lines would get progressively larger over time, which is not the case. Given a fixed water level of 7.0 m at Kratie, the water level at Neak Luong has dropped by about 0.60 m from the pre-dam to mega-dam era. Similarly, at Chaktomuk, the water level dropped by about 0.79 m. Actual measurements of water level reduction due to sand mining are likely to be higher since the incision at Kratie has not been factored in here. Thus, even without decreases in discharge, channel incision would have contributed to the reduction of water levels, leading to the observed decrease in flood pulse in the Cambodian floodplains.

5.4 Wider environmental implications

The Tonle Sap Lake has been decreasing in size throughout the years (Sect. 4.2.2; Kallio and Kummu, 2021; Wang et al., 2020). A parallel can be drawn with Poyang Lake in Yangtze River. There, the Three Gorges Dam has reduced the water level downstream. As the water level was reduced, the hydraulic gradient from Poyang Lake to the Yangtze increased, resulting in a reduction of lake volume there (Zhang et al., 2014, 2015).

In the case of the Tonle Sap Lake, there is a reduction of water levels in the Mekong mainstream during the wet season. As expected from hydrological models (Inomata and Fukami, 2008; Mekong River Commission et al., 2004), the reduced hydraulic gradient from the water levels between the Mekong and the lake leads to lesser water entering the lake from the Mekong. During the pre-dam era, 74.54 km$^3$ of water emptied into the Mekong during the normal flow phase,
while 49.67 km$^3$ of water entered the lake during the reverse flow phase annually, meaning that there was a net outflow of 24.87 km$^3$. In the mega-dam era, only 68.81 km$^3$ of water flowed into the Mekong during the normal phase and reverse flow decreased to 31.74 km$^3$; net outflow from the lake to the Mekong has increased to 31.07 km$^3$. When comparing the difference in discharge towards the Mekong during the pre-dam and mega-dam era, we see that the average annual outflow increased by 6.20 km$^3$. Without a corresponding increase in inflows from either precipitation or lake tributaries, the lake will decrease in volume over time. Based on reconstructed water level data of Guan and Zheng (2021), we estimated the lake volume during 1960–1990 using the volume-stage relation by Kummu et al. (2014). Indeed, compared to the 1960–1990 values, the Tonlé Sap Lake volume during 2010–2019 decreased by 35.9 km$^3$, in line with other studies that observed a shrinking of the lake (Ji et al., 2018; Lin and Qi, 2017; Wang et al., 2020).

Downstream of the Cambodian floodplains, the VMD will also be affected by the decreased flood pulse. While hydropower dam operations can increase dry-season water levels (Dang et al., 2016), the combined effects of local channel incision (Binh et al., 2020b) and irrigation operations will reduce dry-season water. Since irrigation infrastructure in the VMD is even more developed than that in Cambodia (Tran and Weger, 2018), the impacts of dry-season extraction is likely to be even greater in the VMD. Therefore, the decreased flood pulse at Chaktomuk and Neak Luong is likely to propagate further downstream to the VMD. Also, a smaller flood pulse will bring fewer sediments to the VMD (Binh et al., 2021). As a result, the current problems of land subsidence and seawater intrusion may become more severe in the future (Binh et al., 2020a; Kantoush et al., 2017; Zoccarato et al., 2018).

The continued expansion of irrigation and sand-mining operations in the Cambodian floodplains may be unsustainable in the long run. The annual flood pulse is key to regulation of the health of the floodplains, from fisheries to sediment replenishment. For instance, annual floods increase soil health through buffering acidity and increasing its nutrient content (Dang et al., 2016; Sakamoto et al., 2007). A reduction of the flood pulse amplitude and duration will reduce the annual flood extent, thereby reducing soil productivity. Concurrently, as fish migration cycles are intimately tied to the flooding extent, the shift in flood pulse timing will affect catch rates (Baran and Myschowoda, 2009; Enomoto et al., 2011).

### 5.5 Limitations and future directions

Firstly, this study is limited by the hydrological data available. For example, the water level data at Kampong Luang (Tonlé Sap Lake) are only available from 1996 onwards, hampering efforts to understand the hydrology of the lake during the pre-dam era. Secondly, in light of channel incision caused by sand mining and other anthropogenic activities, the stage–discharge relationship at several key stations such as Neak Luong and Chaktomuk should be maintained more frequently. Otherwise, regular bathymetry surveys should be done to assess the rate of channel incision more accurately in the area. Thirdly, given the complex hydrology in the Cambodian floodplains, sensors should be installed along the canals or at key intersections in the floodplains. This way, overland flow – a poorly quantified variable – can be monitored and subsequently be included within more complex hydrological models.

### 6 Conclusion

By using 60 years of data from numerous stations across the floodplains, this study quantified that the flood pulse at the Cambodian floodplains has indeed changed drastically in the past decade. Compared to 1962–1991 levels, minimum water levels in 2010–2019 increased by 0.10 m (Neak Luong) and 1.55 m (Kratie). As a result, the flood amplitude decreased by 5.6 % (Kampong Cham) to 12.9 % (Prek Kdam), meaning that the annual flood extent diminished. Furthermore, the flood season has decreased by 26 d (Kampong Cham) and 40 d (Chaktomuk), with the flood season starting later and ending much earlier.

Correspondingly, the altered flood pulse along the Mekong mainstream affected the annual reverse flows along the Tonlé Sap River. At Prek Kdam, total annual reverse flow dropped from 48.67 km$^3$ in 1962–1972 to 31.74 km$^3$ in 2010–2018, representing a dramatic drop of 56.5 %. Correspondingly, the Tonlé Sap Lake has also been altered by this huge change. There, minimum and maximum water levels have dropped by 0.10 and 1.06 m respectively. These reductions correspond to a decrease in lake area of 3.1 % in the dry season and 20.6 % during the wet season.

These figures were put into perspective as we considered various possible anthropogenic drivers for the hydrological change. We showed that it is unlikely that upstream dams are the main contributing factor for the decline in flood pulse. Instead, the hydrological alterations are more likely to be caused by local factors. The boom in irrigation infrastructure in the last decade has resulted in more water being diverted away from the Mekong and Tonlé Sap Lake to the fields. We estimate that an average of (2.1 ± 0.3) km$^3$ of water was lost per year from the Cambodian floodplains from 1976–2019. Additionally, declining sediments combined with sand-mining operations have caused further channel erosion. Together, irrigation and channel incision has contributed to the observed reduction in flood pulse at the Cambodian floodplains.

As the hydraulic gradient governing the reverse flow to the Tonlé Sap Lake decreases, the lake may suffer a permanent reduction in water volume. Furthermore, the impacts of the runoff reduction will be felt further downstream in the

https://doi.org/10.5194/hess-26-609-2022
VMD. This identified shift in flood pulse is non-trivial, with far-reaching ecological and environmental impacts across international borders. Therefore, policy planners must consider the long-term impact of their plans such that the harvesting of the Mekong can be conducted sustainably.

Data availability. Hydrological data are open source and can be downloaded from the MRC data portal at https://portal.mrcmekong.org/home (Mekong River Commission, 2021). Precipitation data used in this study were also obtained from the MRC portal.

Supplement. The supplement related to this article is available online at: https://doi.org/10.5194/hess-26-609-2022-supplement.

Author contributions. SDXC: conceptualisation, data curation, formal analysis, methodology, writing – original draft, writing – review and editing. LXX: conceptualisation, data curation, methodology, writing – review and editing, funding acquisition. CO: data curation, methodology, writing – review and editing. TS: data curation, methodology, writing – review and editing. CGW: writing – review and editing.

Competing interests. The contact author has declared that neither they nor their co-authors have any competing interests.

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Special issue statement. This article is part of the special issue “Socio-hydrology and transboundary rivers”. It is not associated with a conference.

Acknowledgements. We want to thank the Mekong River Commission (MRC) for access to the hydrological data. We want to thank the Ministry of Water Resources and Meteorology, Cambodia for their support. Also, we would like to thank our three anonymous reviewers and editor Fuqiang Tian for their insightful comments. This study was supported by the National University of Singapore (R-109-000-227-115; R-109-000-273-112).

Financial support. This research has been supported by the National University of Singapore (grant nos. R-109-000-227-115 and R-109-000-273-112).

Review statement. This paper was edited by Fuqiang Tian and reviewed by three anonymous referees.

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