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Mapping of Forest Types in Alaskan Boreal Forests Using SAR Imagery

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Abstract:
Mapping of forest types in the Tanana river flood-plain, interior Alaska, is performed using a maximum-a-posteriori Bayesian classifier applied on SAR data acquired by the NASA/JPL three-frequency polarimetric AIRSAR system on several dates. Five vegetation types are separated, dominated by 1) white spruce, 2) balsam poplar, 3) black spruce, 4) alder/ willow shrubs, and 5) bog/fen/nonforest vegetation. Open water of rivers and lakes is also separated. Accuracy of forest classification is investigated as a function of frequency and polarization of the radar, as
well as the forest seasonal state, which includes winter/frozen, winter/thawed, spring/flooded, spring/ unflooded, and summer/dry conditions. Classifications indicate that C-band is a more useful frequency for separating forest types than L or P-bands, and HV polarization is the most useful polarization at all frequencies. The highest classification accuracy, with 90 percent of forest pixels classified correctly, is obtained by combining L-band HV and C-band HV data acquired in spring as seasonal river flooding recedes and before deciduous tree species have leaves. In 17 forest stands for which actual percentages of each tree species are known, the same radar data are capable of predicting tree species composition with less than 10 percent error. For the same combination of observation channels, classification accuracy is 79 percent in spring on a day of intense river flooding, and 62 percent on a dry summer day with leaves on deciduous trees. In winter, using 4-look SAR data instead of 16-look, classification accuracy is 55 percent on a frozen day, and 76 percent on a thawed day. White spruce and balsam poplar stands are best separated in thawed conditions when balsam poplar trees have no leaves. From our classification, we predict that current and future spaceborne SAR systems will have limited mapping capabilities when used alone. Yet, RADARSAT combined with J-ERS-1 and ERS-1 could resolve forest types with 80 percent accuracy, separate nonforest areas resulting from commercial logging or forest wildfire, and map river edges. For comparison, a combination of green, red, and near-infrared radiance data acquired by SPOT-2 on a dry summer day yields a classification accuracy of 83 percent for the same forest stands, with limited success in distinguishing among deciduous forest types and among coniferous forest types. © 1994 IEEE

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Radar Estimates of Aboveground Biomass in Boreal Forests of Interior Alaska

Eric Rignot, JoBea Way, Cynthia Williams, and Leslie Viereck

Abstract—Airborne SAR data gathered by the NASA/JPL three-frequency, polarimetric, radar system in winter, spring, and summer over the Bonanza Creek Experimental Forest, near Fairbanks, AK, are compared to estimates of whole-tree aboveground dry biomass from 21 forest stands and two clear-cuts. While C-band radar backscatter shows little sensitivity to biomass, L- and P-band radar backscatter increase by more than 6 dB when biomass increases from 5 to 200 tons/ha. Using second-order polynomial regressions, biomass values are predicted from the radar at L- and P-band and compared to actual biomass values. At P-band HV-polarization, the error in predicted biomass is about 30% of the actual biomass. When HV-, HH-, and VV-polarization are used together in the regression, the error in predicted biomass is about 20%. Errors obtained using L-band data are a few percents larger. These errors are caused by uncertainties in actual stand biomass estimates, significant inner-stand spatial variations in biomass, unusual conditions of forest stands following natural disturbances, along with interactions of the radar signals with a complex three-dimensional structure of the canopy. Multiple incidence angle data reveal that the incidence angle $\theta$, of the radar illumination is also a factor influencing the retrieval of biomass, even at HV-polarization, when $\theta > 50^\circ$ or $\theta < 25^\circ$. Finally, the radar response of the forest—and thereby the regression curves for biomass retrieval—are dependent on the seasonal and environmental conditions.

I. INTRODUCTION

Recent experiments using airborne microwave data indicate forest biomass could be quantified remotely [1]-[3]. Radar backscatter was shown to be positively correlated with total aboveground dry biomass until it saturates at a level which is higher with increasing radar wavelengths. The results provide evidence that inversion algorithms could be developed for predicting forest biomass from long wavelength SAR data. At the same time, these studies have been limited to the examination of undisturbed, even-aged, monospecific plantations of pine trees in temperate regions, with nearly level topography. Modeling and simulation results [4]-[7] suggest that the effect on radar backscatter of tree structure, vegetation moisture condition and status, as well as understory condition and topography can be significant and thereby complicate the inference of aboveground biomass from the radar. Before operational applications are developed it is necessary to investigate experimentally the relative magnitude of these effects in natural forest ecosystems.

Here we examine the relationship between radar backscatter and aboveground biomass in a natural forest setting, with mixtures of coniferous and deciduous tree species, strongly varying environmental conditions, and level to moderate topography. The forest site is the Bonanza Creek Experimental Forest (BCEF), a Long Term Ecological Research (LTER) site located at 64° 45'N, -148°W, in the proximity of Fairbanks, AK, along the Tanana river. SAR data were gathered at BCEF on several occasions and at different seasons by the three-frequency NASA/JPL polarimetric airborne radar system (AIRSAR). BCEF is in the midst of the boreal forested biome which is believed to play an important role in the global carbon cycle [8]-[12]. Boreal forests are a major reservoir of live carbon, only second to broadleaf humid forests, and are thought to be responsible for the seasonal variations in atmospheric CO$_2$ observed at the northern latitudes. Mapping and monitoring of whole-tree aboveground biomass in these forests is therefore particularly relevant to studies of the global carbon cycle.

II. METHODS

A. Study Area

The vegetation landscape at BCEF is a mosaic of forest, grassland, shrubs, bog, and tundra types that have formed primarily as a result of slope, aspect, elevation, parent material, and succession after wildfire [13]-[15] (Fig. 1). Presence or absence of permafrost, and solar illumination angle, both correlated with slope and aspect, are dominant factors in the distribution of vegetation types. The forest site includes both upland fire-controlled succession, and floodplain succession forests. Upland forest types vary from highly productive aspen (Populus tremuloides), paper birch (Betula papyrifera), and white spruce (Picea glauca) stands on well-drained, nutrient-poor, permafrost soils of north-facing slopes, lowlands, and lower slopes. In mature white spruce stands in the uplands, dominant tree sizes average about 30 cm DBH (diameter at breast height) and 25-35 m in height, with tree densities about 500-1000 stems/ha. Standing aboveground tree biomass ranges from 92 to 183 tons/ha for the hardwood stands up to 249 tons/ha for white spruce stands. In mature, black spruce, upland stands, dominant tree sizes average about 6-15 cm, and 5-10 m in height, with tree densities about 1400-4000 stems/ha. Aboveground biomass ranges from 2 to 11 tons/ha.

Floodplain forests vary from productive stands of balsam poplar (Populus balsamifera) and white spruce forming on river alluvium, permafrost-free soils, to slow-growing black spruce and bog occupying older terraces underlain by permafrost. Young successional stages are dominated by alder (Alnus tenuifolia) and willow shrubs (Salix spp.), followed by extensive stands of balsam poplar and the mature stages of well-developed stands of white spruce. In mature, white spruce stands, dominant tree sizes average about 30 cm DBH and 25-35 m in height, and tree densities are high (1000 stems/ha). Aboveground tree biomass ranges from 40-180 tons/ha for balsam poplar stands to 60-245 tons/ha for white spruce stands. Black spruce is the most widespread forest type, covering 44% of...
Fig. 1. Composite cross-section showing the distribution of forest vegetation types with topography in the Tanana uplands and lowlands in the Fairbanks vicinity. Whole-tree above-ground dry biomass is presented for selected vegetation types [7].

The climate at BCEF is continental, with large diurnal temperature variations, low precipitation, low cloud cover and low humidity [16]. Mean annual temperature is \(-3.5^\circ\text{C}\), and temperature extremes range from \(+35^\circ\text{C}\) in June to \(-65^\circ\text{C}\) in January. The annual mean precipitation is 286 mm, with 30% falling as snow.

B. Airborne SAR Data

AIRSAR is a polarimetric radar imager operating at P- (68 cm wavelength), L- (24 cm) and C-band (5.6 cm) frequencies simultaneously [17]. AIRSAR imaged BCEF on March 13, 17, and 19 of 1988 when the forest changed from frozen (March 11) to thawed conditions (March 13) and back to frozen conditions (March 17) due to a transition to unusually warm air-temperatures for the season [18]. On May 4, 6, and 7 of 1991, AIRSAR overflew the study site when the forest changed from flooded (May 4) to unflooded conditions (May 6–7) with the formation and subsequent disappearance of ice jams in the Tanana river. Finally, AIRSAR overflew BCEF in the Summer of 1993, with clear skies and dry climatic conditions. Data calibration was performed in a similar fashion for all data takes using the calibration procedure described in [19] and 1.8 m trihedral corner reflectors deployed on the ground in clearcut areas prior to each of the SAR overflights. The SAR data were subsequently projected onto ground-range, resampled to a common pixel spacing, and co-registered to the scene acquired on May 6, 1991 using the registration procedure for polarimetric SAR data described in [20]. All data takes considered in this study were acquired at a look angle of 40° unless mentioned otherwise.

C. Ground Data Collection

Between 1988 and 1993, 21 forest stands were inventoried for tree density, DBH, height, canopy depth, and species. Aerial photographs were used to determine the starting point and transect direction for each stand. The first plot was started well enough within the stand to avoid any edge effects from the river, adjacent clearcuts, and other stands. Ten plots were sampled in each stand, separated by about 100 m, eventually reduced to 50 m in smaller stands. Transects turned in many instances to stay within stand boundaries and to avoid anomalies such as old river sloughs. Plot radii were chosen to include roughly 15 live trees (a "live" tree has a DBH > 2.5 cm) providing a sampling of 150 trees per stand. While all DBH's were measured, not all trees were measured for tree height. In each stand, only 40 trees of each species present in the overstory were chosen for height measurements. From the DBH and height data collected in the field, linear regressions relating DBH and height of measured trees were derived and then used to estimate heights of the trees for which height had not been measured. Whole-tree aboveground biomass was then estimated for each tree using empirical, allometric equations from Manning et al. [21], Singh [22], and Yarie and Van Cleve [23] for all tree species except alder for which we used a linear regression analysis developed by the Forest Soils Laboratory, University of Alaska, Fairbanks. Biomass values for each tree were subsequently averaged over each plot, and biomass values for the plots were used to compute the average and standard deviation of the stand biomass. The inventory results are summarized in Table I.

The transect boundaries of each stand were drawn on the computer to within a few SAR pixels (about 30 m) of their actual position recorded on aerial photos. Calibrated radar backscatter values, \(\sigma^a\), were then extracted from each stand at each frequency and polarization and averaged together over the whole stand. Even for the smallest stands, several hundred independent SAR samples were averaged together so \(\sigma^a\) is not affected by image speckle. To compute the standard deviation in \(\sigma^a\), we used the statistical model described in [24] to remove the contribution from speckle (1991 and 1993 AIRSAR data are 16-looks; 1988 AIRSAR data are...
TABLE I

STATIC CHARACTERISTICS OF 21 FOREST STANDS FROM BCEF, ALASKA. WS IS WHITE SPRUCE, BP IS BALSAM POPLAR, BS IS BLACK SPRUCE, BP IS BALSAM POPLAR. AL IS ALDER, BIR IS BIRCH, TA IS TAMARACK, AND WI IS WILLOW. TREE SPECIES COMPOSITION ARE SHOWN IN % BASED ON THE NUMBER OF TREES PER SPECIES PER STAND. WS40, WS41, WS42, AND PBW40 ARE FROM THE UPLANDS. ALL OTHER STANDS ARE FROM THE FLOODPLAINS. IN THE FIRST COLUMN, (yr) INDICATES THE YEAR THE STAND WAS SAMPLED. ANGLE IS THE LOCAL INCIDENCE ANGLE OF THE RADAR ILLUMINATION COMPUTED AT THE CENTER OF THE FOREST STAND FOR THE SCENE ACQUIRED ON MAY 6, 1991.

| STAND | (yr) | Stem #/ha | Height (m) | Age yr | DBH cm | Sem. B tons/ha | Wt. B tons/ha | Tree Species % | Angle (deg) |
|-------|------|-----------|------------|--------|--------|---------------|--------------|----------------|-------------|
| WS1   | 89   | 144       | 22.1       | 50.0   | 165    | 19.6          | 213 (44)     | 217 (48)       | 35.9        |
| WS2   | 68   | 507       | 25.1       | 64.0   | 160    | 16.5          | 157 (39)     | 167 (46)       | 33.6        |
| WS3   | 69   | 600       | 18.8       | 124    | 19.8   | 117 (28)      | 173 (38)     | 178 (46)       | 35.6        |
| WS4   | 60   | 608       | 22.0       | 125    | 22.0   | 153 (45)      | 163 (55)     | 163 (55)       | 37.6        |
| WS5   | 61   | 377       | 21.3       | 177    | 17.9   | 181 (41)      | 181 (46)     | 151 (48)       | 35.9        |
| WS6   | 67   | 1175      | 21.5       | 120    | 21.2   | 221 (65)      | 221 (65)     | 217 (64)       | 32.4        |
| WS7   | 66   | 1330      | 23.9       | 142    | 13.9   | 209 (53)      | 209 (53)     | 209 (53)       | 32.2        |
| WS8   | 66   | 1500      | 19.4       | 118    | 21.8   | 156 (43)      | 156 (48)     | 156 (48)       | 33.9        |
| WS9   | 65   | 353       | 20.0       | 132    | 24.0   | 109 (46)      | 109 (46)     | 109 (46)       | 50.1        |
| WS10  | 60   | 497       | 19.1       | 50     | 8.3    | 110 (32)      | 109 (32)     | 109 (32)       | 42.4        |
| WS11  | 67   | 1060      | 19.5       | 75.0   | 15.4   | 137 (62)      | 136 (62)     | 135 (62)       | 40.6        |
| BS1   | 69   | 1975      | 7.6        | 200    | 8.8    | 37 (8)        | 37 (8)       | 37 (8)         | 35.9        |
| BS2   | 60   | 1407      | 6.6        | 200    | 6.8    | 22 (5)        | 22 (5)       | 22 (5)         | 41.1        |
| BS4   | 61   | 1467      | 6.3        | 87     | 4.4    | 8 (4)         | 8 (4)        | 8 (4)          | 51.2        |
| BS5   | 61   | 1615      | 10.8       | 75     | 17.0   | 192 (109)     | 179 (107)    | 179 (107)      | 38.9        |
| BS6   | 60   | 1235      | 11.3       | 75     | 15.2   | 196 (99)      | 196 (99)     | 196 (99)       | 38.8        |
| BS8   | 65   | 1528      | 12.5       | NA     | 13.3   | 85 (47)       | 85 (49)      | 85 (49)        | 41.5        |
| BS9   | 60   | 939       | 7.1        | 30     | 7.3    | 32 (11)       | 32 (11)      | 32 (11)        | 44.1        |
| AL1   | 60   | 760       | 6.9        | 30     | 9.2    | 22 (10)       | 22 (10)      | 21 (10)        | 38.8        |
| PBW40 | 65   | 471       | 19.4       | NA     | 10.7   | 165 (45)      | 165 (45)     | 165 (45)       | 57.4        |

4-looks) and obtain a measure of the natural spatial variability in \( \sigma^0 \) or "texture" of the forest.

III. RESULTS

The calibrated radar response \( \sigma^0 \) of the forest on May 6, 1991 is shown in Fig. 2 as a function of whole-tree aboveground dry biomass. The data acquired in winter and summer are discussed later on in the paper. As \( \sigma^0 \) eventually approaches a saturation level at the high biomass levels, third-order polynomials were fit to the data instead of straight lines (Fig. 2). The results show that \( \sigma^0 \) is positively correlated with aboveground biomass. In contrast, the normalized difference vegetation index (NDVI) of the same forest stands computed from a SPOT scene acquired on August 8, 1991 [25] is not correlated with biomass (Fig. 3). Although NDVI could be used to distinguish deciduous species (alder and balsam poplar with higher NDVI) from coniferous species (black spruce and white spruce with lower NDVI), NDVI does not provide indications on the relative biomass level of each forest stand.

At C-band [Fig. 2(a)], \( \sigma_{\text{AV}} \) increases with biomass until it satu-
the enhanced sensitivity of H-polarization to biomass. The dynamic range in the logarithm of biomass to second-order polynomials in the logarithm of \( \sigma_{HV} \), \( \sigma_{HH} \), or \( \sigma_{VH} \) at one frequency, i.e., of the form: 

\[
\log B = \sum_{k=0}^{2} a_k \sigma_{HV}^k
\]

where \( B \) is the stand biomass, \( \sigma_{HV} \) is expressed in decibels, \( X \)'s and \( Y \)'s are \( H \)'s or \( V \)'s, and \( a_k \)'s are the coefficients of the polynomial. Biomass is then predicted from the radar measurements and compared to inventory estimates (Fig. 4).

At L-band HV, the radar largely overestimates the biomass of alder stands. When alder stands are included, the error rate is 75%, whereas when alder stands are ignored it is 39%. One reason is that \( \sigma_{HH} \) of alder trees at L-band is higher than expected from their relatively low biomass level, and we use the same inversion curve for all tree species. For clear-cuts, black spruce, and young stands of white spruce the radar estimates are accurate below a 100 tons/ha. Above that level, data dispersion increases. The results improve slightly when \( \sigma_{HH} \) and \( \sigma_{VH} \) are also used in the regression as the error rate drops by 4% (Table II).

Radar predictions at P-band HV are more accurate than at L-band HV and extend over almost the entire range of biomass level for these forests [Fig. 4(c)]. The error rate without alder trees is 29% (Table II), and 54% when including alder trees. Adding \( \sigma_{HH} \) and \( \sigma_{VH} \) to the regression significantly improves the results. The error rate drops to 21% without alder trees and 33% otherwise. The biomass level of alder trees is more correctly estimated. One reason is that \( \sigma_{HH} \) < \( \sigma_{VH} \) for alder trees whereas at the high biomass levels \( \sigma_{HH} >> \sigma_{VH} \). Indeed, it is expected that trunk-ground interactions would be low in magnitude in alder stands compared to volume scattering from the branches because alder trees are typically short (Table I). One exception is BP6 for which the predicted biomass is too high. Yet, BP6 is not a pure stand of alder as it contains a large proportion of tall balsam poplar trees which can yield strong trunk-ground interactions. BP6 was also flooded on May 4 and probably still wet on May 6, which increases the likelihood of observing higher than usual trunk-ground interactions in that stand. Similarly, large errors are found in other stands exhibiting particular conditions. For instance, BP14 is a very heterogeneous stand with a large uncertainty in actual biomass (Table I) and the radar does not accurately predict biomass in that stand. The radar also overpredicts forest biomass in WSBP2, a stand known for exhibiting enhanced radar backscatter in all radar data taken compared to other stands of balsam poplar. We recently found out that most balsam poplar trees within the stand are rotten, with very high dielectric constant values for the tree trunks. A higher than usual water content for those trees may be the reason for the enhanced radar backscatter from the tree-trunk observed in WSBP2 and an overestimation of biomass.

A similar inversion was performed using the radar data acquired on July 21, 1993. However, only 19 stands could be used in the
Fig. 4. Predicted biomass levels of various forest stands from BCEF versus actual biomass levels at (a) L-band HV-polarization; (b) L-band HH-, HV-, and VV-polarizations; (c) P-band HV-polarization; and (d) P-band HH-, HV-, and VV-polarizations.

TABLE II

| Frequency - Polarization(s) | Error on May 6, 1993 | Error on July 21, 1993 |
|-----------------------------|----------------------|------------------------|
| P-band HV                   | 29                   | 32                     |
| P-band HH, HV, VV           | 21                   | 19                     |
| L-band HV                   | 39                   | 27                     |
| L-band HH, HV, VV           | 35                   | 26                     |

Analysis because WS41 and WS40 were not imaged in 1993 due to a reduction in swath width of the radar; and AL1 and BP6 suffered serious damage in 1992, leaving all alder trees permanently broken or laying flat on the ground. Hence, the summer estimates do not include alder stands (Table II). Nevertheless, the results are comparable to those obtained for the spring data, perhaps even slightly better in the summer. The major difference is that the inversion curves are different between the two dates (Section V).

For comparison, we performed a similar regression on the radar data collected over the Landes forest plantation, France [1]. The Landes forest plantation is more appropriate for biomass retrieval because trees are mono-species, very uniform in density, height, DBH, and age within each stand due to seeding and management practices. Stand biomass is known with 12.5% accuracy and is less than 150 tons/ha. The results show little data spread and an excellent retrieval of forest biomass from the radar (Fig. 5). The error rate is 11% at P-band whether one or three polarizations are used, 13% at L-band HV, and 20% at L-band using all three polarizations. In those areas of moderate biomass, \( \sigma_\text{u} \) and \( \sigma_\text{v} \) do not help improve the estimates obtained using only \( \sigma_\text{u} \). It was understood that in natural forest ecosystems the accuracy of the inversion will be less because tree density, height, diameter and age as well as canopy structure have large spatial variations. At BCEF, standard deviation in biomass within each stand is 34% of the average stand biomass (Table I) and as high as 79% for BP14. Tree height, density, and height commonly vary by more than one order of magnitude from one plot to the next or even within the same plot. Hence, uncertainties in stand biomass increase with biomass and are very large for mature, tall forests. Actual uncertainties may even be larger considering that we ignore uncertainties associated with the use of empirical, allometric equations for computing tree biomass. Another complicating factor is that radar signals interact with more complex tree canopies of varying three-dimensional structure because the forest stands are not spatially homogeneous,
V. TEMPORAL DYNAMICS

The radar backscatter from the forest is dependent on the seasonal and environmental conditions. As an illustration, Fig. 6 shows $\sigma^0$ from the forest on March 17, 1988 when the forest was frozen, and on July 21, 1993 when the forest was thawed and dry, with fully developed canopies. C-band data are not shown because they are less relevant to the retrieval of total aboveground biomass. A complete discussion of the radar scattering characteristics of the forest in the summer will be reported in a future paper. P-band data from March 17, 1988 are also not shown because they could not be absolutely calibrated with confidence. Fig. 6(a) shows that $\sigma^0$ at L-band may change by as much as 6 dB between spring and winter/frozen, which is as much as the sensitivity of $\sigma^0_{HV}$ to stand biomass. An inversion curve derived from winter/frozen data could not yield reliable estimates of biomass in spring. Similarly, $\sigma^0$ from the forest changes by several dB between spring [Fig. 2(b)] and summer [Fig. 6(b)], depending on the polarization, and regression curves for biomass retrieval again differ significantly between the two dates. At P-band, $\sigma^0$ from the forest in the summer [Fig. 6(c)] is several decibels lower than in spring [Fig. 2(c)].

Radar mapping of aboveground biomass is therefore dependent on seasonal and environmental conditions. At the same time, the temporal variability of $\sigma^0$ provides additional information about the forest. For instance between March 13 and March 17, 1988, the measured 6 dB change in $\sigma^0$ of the forest was due to a dramatic change in the dielectric constant of the tree trunks associated with the freezing of the fresh liquid water content of the trees on March 17 [18]. Hence, on March 17, the radar signals were sensitive to the frozen biomass, while on March 13, 1988 the radar signals were sensitive to the moist biomass of the forest. In the summer, the radar returns from the forest are lower presumably because the forest canopy and ground layers were drier than in early spring, and the radar signals were more attenuated during propagation through the canopy due to the presence of leaves on deciduous trees.

VI. INFLUENCE OF THE INCIDENCE ANGLE

Another factor influencing the retrieval of forest biomass from the radar is the incidence angle of the radar illumination onto the forest. On May 6, 1991, AIRSAR imaged BCEF successively at L-band and P-band.

Fig. 6. $\sigma^0$ values of various forest stands versus whole-tree aboveground dry biomass (tons/ha) for SAR data acquired on (a) March 17, 1988 at L-band; (b) July 21, 1993 at L-band; and (c) July 21, 1993 at P-band.

include tree gaps, old river sloughs, and a mixture of different tree species of various ages.
Comparison of the dependence of radar illumination for multiple incidence angle SAR data acquired on May 6, 1991 flight, and for which \( \theta_i = 53^\circ, 61^\circ, \) and \( 58^\circ \). These stands, labeled PBW40 (Table I), BS21 (not listed), and PB41 (not listed) have a stand biomass of 103, 24, and 92 tons/ha, but the radar estimates are, respectively, 42, 4, and 16 tons/ha, i.e., significantly lower than the actual values. Hence, when \( \theta_i < 25^\circ \) or \( \theta_i > 50^\circ \) or \( \theta_i = 58^\circ \) must be included in the inversion. A direct consequence is that topographic information will in general be required for reliable mapping of aboveground biomass over large areas.

VII. CONCLUSIONS

This study suggests SARs operating at long radar wavelengths have potential for mapping aboveground biomass of boreal forests in interior Alaska, a variable of prime importance to carbon cycling investigations, ecosystem studies and forest management applications. The error in predicted biomass from the radar is about 20% of the actual biomass at P-band. L-band signals do not perform as well because of more pronounced dependence on tree species (for instance alder trees), but the difference in performance between the two frequencies remains small. Prediction of forest biomass from the radar in monospecie forest plantations is 10% better, but this was an expected result. Data dispersion is more significant for natural forest ecosystems than for forest plantations because uncertainties in estimating aboveground biomass are larger in natural forests due to spatial variations in tree species, age, density, height, and DBH, and also because the radar signals interact with a canopy of spatially varying three-dimensional structure. Several other factors are also of importance for developing operational applications. One is that changing environmental conditions have a pronounced effect on radar backscatter from the forest, especially at the high southern latitudes where drastic changes occur. Hence, inversion curves derived at one geographical location under different environmental conditions may be different. Secondly, digital elevation models should be combined with the radar data to improve radar estimates of forest biomass in areas of moderate topography.

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Abstract—This communication describes a procedure to estimate the vertical profiles of stratospheric aerosols based on the solar occultation measurements of the improved limb atmospheric spectrometer (ILAS), which will be installed on the advanced earth observing satellite (ADEOS) in 1996. It is well known that stratospheric aerosols play a dominant role in global climatic change [1]. Stratospheric aerosols are also important for the formation of high altitude clouds, such as the polar stratospheric clouds, and as tracers to study the dynamics of the stratosphere. For this reason, global-scale measurements of stratospheric aerosols are desirable [2]-[4].

ILAS is designed to measure vertical profiles of polar stratospheric trace gas, temperature, pressure, and aerosols by using two observing channels at the visible and infrared wavelengths covering 0.753 − 0.784 µm and 6.2 − 11.8 µm, respectively.

An inversion algorithm is a key process for our aerosol retrieval method [5]. The vertical structure inversion provides profiles of local atmospheric extinction from ILAS transmittance data. This is followed by a species separation step that decomposes the obtained total extinction into each component of the atmospheric constituents, i.e., air molecules and aerosols. It is shown that the retrieved pressure profiles from ILAS oxygen A-band analysis are available to improve and simplify the aerosol retrieval [6].

II. RETRIEVAL ALGORITHMS

ILAS measures the mean transmittance $T_{\lambda}$ of the atmosphere at wavelength $\lambda$ and tangent height $h$.

\[ T_{\lambda}(h) = \exp \left[ -r_{\lambda}(h) \right] = \exp \left[ - \int_0^h \sigma_{\lambda}(l) \, dl \right]. \]  \hfill (1)

\[ \sigma_{\lambda} = \sigma_{\lambda}^{\text{Ray}} + \sigma_{\lambda}^{\text{D}} + \sigma_{\lambda}^{\text{geo}} \] \hfill (2)

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