Proper Interpretation of Dissolved Nitrous Oxide Isotopes, Production Pathways, and Emissions Requires a Modelling Approach

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Abstract

Stable isotopes (δ¹⁵N and δ¹⁸O) of the greenhouse gas N₂O provide information about the sources and processes leading to N₂O production and emission from aquatic ecosystems to the atmosphere. In turn, this describes the fate of nitrogen in the aquatic environment since N₂O is an obligate intermediate of denitrification and can be a by-product of nitrification. However, due to exchange with the atmosphere, the δ values at typical concentrations in aquatic ecosystems differ significantly from both the source of N₂O and the N₂O emitted to the atmosphere. A dynamic model, SIDNO, was developed to explore the relationship between the isotopic ratios of N₂O, N₂O source, and the emitted N₂O. If the N₂O production rate or isotopic ratios vary, then the N₂O concentration and isotopic ratios may vary or be constant, not necessarily concomitantly, depending on the synchronicity of production rate and source isotopic ratios. Thus prima facie interpretation of patterns in dissolved N₂O concentrations and isotopic ratios is difficult. The dynamic model may be used to correctly interpret diel field data and allows for the estimation of the gas exchange coefficient, N₂O production rate, and the production-weighted δ values of the N₂O source in aquatic ecosystems. Combining field data with these modelling efforts allows this critical piece of nitrogen cycling and N₂O flux to the atmosphere to be assessed.

Introduction

Nitrous oxide (N₂O) is a powerful greenhouse gas, 298 times more potent than CO₂ over a 100-year time line [1]. Atmospheric N₂O concentrations have been increasing at a rate of 0.25%/year over the last 150 years [2]. Consequently, the global N₂O budget has been the subject of intensive research efforts over the past few decades. N₂O is produced through multiple microbial pathways: hydroxylamine oxidation during nitrification and as an obligate intermediate during denitrification and nitrifier–denitrification. Because these pathways of N₂O production have different stable isotopic enrichment factors, isotopic analysis of N₂O can potentially distinguish N₂O produced through different pathways or from different sources [3]. Identifying N₂O sources will provide insights on the fate of N at the ecosystem-scale (e.g., [4–6]). The isotopic ratios of N₂O produced in soil environments (e.g., [7–11]), and in aquatic environments (e.g., [12–18]) have been measured to some extent. Although N₂O production in rivers and estuaries is a significant portion of the global N₂O budget [approximately 1.5 TgN/year, [19]], few studies report isotopic data for rivers [5,20,21].

In ice-free aquatic ecosystems, the δ¹⁵N and δ¹⁸O of dissolved N₂O is affected by gas exchange with the atmosphere. As a result, the isotopic ratios of dissolved N₂O are not equal to those of the N₂O produced within the aquatic ecosystem and continue to change as atmospheric exchange (both ingassing and outgassing) occurs. In addition, isotopic fractionation during influx and efflux causes the isotopic ratios of N₂O flux emitted to the atmosphere to be different than that of the dissolved N₂O [22]. Thus, the simple method of calculating the instantaneous isotopic ratios of the N₂O flux by taking measured dissolved isotopic ratios, adding an equilibrium isotope fractionation, and applying them to measured flux rates is inappropriate. Adjustments of measured isotopic ratios are necessary to understand the isotopic ratios of both produced and emitted N₂O.

In this paper, we present a dynamic model of the stable isotopic composition of both the dissolved and emitted N₂O in aquatic systems. We apply this model to two different measured diel patterns of the isotopic ratios of N₂O in an aquatic ecosystem. We use the model to elucidate the relationship between the isotopic ratios of source, dissolved, and emitted N₂O, to allow for improved interpretation of dissolved N₂O isotope data. Ultimately, a process-based understanding on N cycling with aquatic ecosystems may be developed based on interpretation of N cycling processes.
Materials and Methods

Stable Isotopes of N₂O

N₂O is an asymmetric molecule: the most abundant isotopologues of N₂O are [14N14N18O], [15N14N16N18O], [16N15N18O] and [14N15N18O]. The isotopic ratios, ¹⁵N:¹⁴N and ¹⁸O:¹⁶O, are:

\[
15R = \frac{[15N14N16O] + [14N15N18O]}{2[14N14N16O]} = \frac{[15N2O]}{[14N2O]}
\]

\[
18R = \frac{[14N18N18O]}{[14N14N16O]} = \frac{N_{18O}}{N_{16O}}
\]

where [14N14N16O], [15N14N16O], [14N15N18O] and [14N14N18O] represent the concentrations of the various N₂O isotopologues. Note that 15R is the bulk ¹⁵N:¹⁴N ratio and represents an average ratio of the two ¹⁵N isotopomers and isotopic ratios are reported as δ¹⁵N relative to air and δ¹⁸O relative to VSMOW. Although the isotopic ratio of the ¹⁵N isotopomers can be measured (e.g., [23–25]), the gas exchange fractionation factors are not affected by the intramolecular distribution of ¹⁵N [22]. Many laboratories cannot measure the intramolecular distribution of ¹⁵N and analysis of the bulk ¹⁵N:¹⁴N ratio of N₂O is more common [26]. Here, we confine our analysis to bulk ¹⁵N:¹⁴N ratios and use ¹⁵N₂O to represent the average abundance of the two ¹⁵N isotopomers. The same approach could easily be extended to consider each isotopologue separately.

Dynamic Isotope Model for Dissolved N₂O

A simple three box model (SIDNO, Stable Isotopes of Dissolved Nitrous Oxide) was created using Stella modelling software (version 9.1.4, http://www.sesystems.com) in order to study the relationships between the isotopic ratios of source, dissolved and emitted N₂O (model file is available at https://github.com/jvrenk/SIDNO and by contacting the corresponding author). This model is an adaptation of the isotopic gas exchange portion of the PoRGy model [27], which successfully modelled diel isotopic ratios of O₂ resulting from photosynthesis, respiration, and gas exchange in aquatic ecosystems. One key difference is photo-synthetically produced O₂ in PoRGy has a δ¹⁸O value fixed by the H₂O molecules, whereas SIDNO has N₂O production δ¹⁵N and δ¹⁸O values that can vary independently of each other and of N₂O production rate in order to simulate variability in nitrification and denitrification.

One box in SIDNO is used for the total mass of dissolved N₂O and two additional boxes for the dissolved masses of the two heavy isotopologues (¹⁵N₂O and N₂O). The boxes are open to the atmosphere for gas exchange, are depth agnostic, and each box can gain N₂O via a production term; there is no N₂O consumption term since the δ values of N₂O are largely controlled by the production pathways [28,29] though certain waters can exhibit significant N₂O reduction to N₂ [30,31]. The masses and magnitude of the flows of ¹⁵N₂O and N₂O relative to bulk N₂O are used to calculate the isotopic composition of source, dissolved, and emitted N₂O. Although isotopic ratios are used in the model, we discuss δ values that are common for reporting isotopic ratios. N₂O production rate and its δ values are user-defined and can be adjusted for diel patterns in N₂O production that may be caused by variable O₂ levels [32–37].

Stable Isotope Dynamics of Gas Exchange

The δ values of the net gas exchange flux are controlled by the kinetic fractionation factors for evasion (Rgas/Rdissolved) and invasion (Rinvaded/Rgas). These two δ values are related to the equilibrium fractionation factor: δgas = δRgas and δdissolved = δRdissolved. Therefore, at equilibrium, dissolved N₂O has δ values slightly greater than these at 7.48‰ and 43.75‰, respectively.

In the model, net N₂O flux between the atmosphere and dissolved phase was calculated using the thin boundary layer approach as:

\[
\text{Flux} = k \left( p_{N_2O} \times k_H - [N_2O]_{dissolved} \right)
\]

where the N₂O flux is calculated in mol/m²/h, k is the user-modifiable gas exchange coefficient (m/h), the partial pressure of tropospheric N₂O (assumed to be 320 ppb) after data provided by the ALE GAGE AGAGE investigators, [39,40], kH is the Henry constant for N₂O (mol/atm-m³), and [N₂O]dissolved is the dissolved concentration of N₂O (mol/m³). kH is a function of water temperature [41]:

\[
k_H = 0.025e^{-2600 \left( \frac{T}{298.15} \right)}
\]

where T is temperature in kelvins.

Gas exchange is a two-way process. The net N₂O flux rate (the difference between the invasion and evasion rates) depends on the dissolved N₂O concentration. When a solution is at equilibrium with the atmosphere, the invasion and evasion rates will be equal, and the net flux will be zero.

As with the bulk N₂O flux, the flux of the heavy isotopologues (¹⁵N₂O and N₂O) can be calculated by including the kinetic fractionation factors for N₂O (adapted from [27]):

\[
\text{Flux}^{15}N_2O = k \left( z_{15}^{15}N_2O \times k_H - z_{15}^{15}N_2O_{dissolved} \right)
\]

\[
\text{Flux}N_2^{18}O = k \left( z_{18}^{18}N_2O \times k_H - z_{18}^{18}N_2O_{dissolved} \right)
\]

where p¹⁵N₂O and p¹⁸N₂O are the partial pressures of ¹⁵N₂O and ¹⁸N₂O.

Results

Test of Model Performance

To test the ability of SIDNO to reproduce observed isotopic data, input parameters (N₂O production rate, N₂O δ values, and k) were set to replicate a series of experiments designed to derive fractionation factors for N₂O gas exchange [22]. In these experiments, degassed water was exposed to N₂O gas of known isotopic ratios in a sealed container to varying degrees of saturation.
Modelled dissolved N\(_2\)O concentration and \(\delta\) values increased in response to gas exchange (Figure 1). The model fit to the experimental data is comparable to the original best-fit derivations \((R^2 = 0.77\) for \(\delta^{15}\)N and \(R^2 = 0.82\) for \(\delta^{18}\)O for both the original fit [22] and the SIDNO fit) (Figure 1). The initial isotopic composition of dissolved N\(_2\)O was identical to the gas phase \(\delta^{15}\)N value, but the \(\delta^{18}\)O of dissolved N\(_2\)O was slightly less than the gas phase \(\delta^{18}\)O value. Ultimately, at 100% saturation the \(\delta\) values of the dissolved N\(_2\)O were greater than those of the gas phase as a result of \(\Delta q\). The model successfully simulated the kinetic and equilibrium fractionations during gas exchange under the experimental conditions.

Figure 1. Comparing the model output to the experimental data of [22]. The coefficient of determination for experimental data and SIDNO model outputs were comparable to those of [22], \(R^2 = 0.77\) for \(\delta^{15}\)N and \(R^2 = 0.82\) for \(\delta^{18}\)O. Precision of measurements for the experimental data was \(\pm 0.05\%\) for \(\delta^{15}\)N and \(\pm 0.1\%\) for \(\delta^{18}\)O. doi:10.1371/journal.pone.0090641.g001

Next, SIDNO was used to provide insight into the effect of degassing on the \(\delta\) values of dissolved and emitted N\(_2\)O. Here the results of two model runs with the same initial N\(_2\)O concentration but different initial \(\delta\) values of dissolved N\(_2\)O were compared (Figure 2). As N\(_2\)O saturation declined both the dissolved \(\delta^{15}\)N values and instantaneous \(\delta^{15}\)N values of the emitted N\(_2\)O remained relatively constant, dissolved \(\delta^{18}\)O values and instantaneous \(\delta^{15}\)O values of the emitted N\(_2\)O varied by about 10, when the solution was very supersaturated (\(>300\%\) saturation). The \(\delta\) values rose quickly as the system approached 100% saturation. Because the light isotopologue diffuses out of solution faster than the heavy isotopologue, the instantaneous \(\delta\) values of the emitted N\(_2\)O were always less than the concomitant \(\delta\) values of dissolved N\(_2\)O. The isotopologues of N\(_2\)O reached equilibrium independently of each other and therefore the total mass emitted for each isotopologue and rate of change depended on the initial concentration and \(\delta\) values. The retention of N\(_2\)O in the dissolved phase caused the \(\delta\) values of the mass emitted to differ from those of total mass production. However, when initial dissolved N\(_2\)O concentrations were high (\(>1000\%\) saturation) the \(\delta\) values of the total N\(_2\)O emitted were similar to the \(\delta\) values of dissolved N\(_2\)O because the mass of N\(_2\)O lost is very much larger than the N\(_2\)O that remained dissolved. The value of \(k\) did not affect the gas exchange trajectories only the speed at which the system reached equilibrium.
N$_2$O isotope data are often plotted as $\delta^{18}$O versus $\delta^{15}$N to elucidate relationships between the various sources and tropospheric N$_2$O [38]. The trajectories on these plots (Figure 2C) were dictated by the $\delta$ values of the source relative to the constant atmospheric value and the $\alpha$ values. Note that some plots in the literature differ due to different reference materials for the $\delta^{18}$O scale (VSMOW and atmospheric O$_2$).

**Modelling Scenarios with Steady State Production of N$_2$O**

The SIDNO model can be used to probe the stable isotope dynamics of N$_2$O in a variety of situations that may be encountered in aquatic environments to elucidate the relationship between the N$_2$O source (a function of N cycling processes), dissolved (the easily measured component), and emitted (of consequence for greenhouse gas production and global N and N$_2$O cycle).

In the steady-state production of N$_2$O (constant rate and $\delta$ values), by definition, the $\delta$ values of N$_2$O production must be the same as those of the emitted N$_2$O. As a result, the $\delta$ values of the dissolved N$_2$O cannot equal that of the source (or emitted) N$_2$O at steady state because the dissolved N$_2$O must be offset from the emitted N$_2$O by at least the $\alpha$ values. As the steady-state production rate was increased, the steady-state N$_2$O concentration increased and the dissolved $\delta$ values approached but did not equal the source (Figure 3). Even at moderate supersaturations ($\leq 1000\%$) the effect of atmospheric N$_2$O equilibration on the $\delta$ values of dissolved N$_2$O cannot be ignored.

At steady state, the $\delta$ values of the emitted N$_2$O must be equal to the source; the large difference between source/ emitted and dissolved N$_2$O underscores the importance of adjusting the measured $\delta$ values of dissolved N$_2$O in order to determine aquatic contributions of N$_2$O to the atmosphere or N$_2$O sources. This is critical when using dissolved measurements of N$_2$O to constrain the global isotopic N$_2$O budget, but not been done in most studies, e.g., [16,42–44] but see [45].

**Modelling Scenarios with Variable Production of N$_2$O**

The relationship between the $\delta$ values of source, dissolved, and emitted N$_2$O are much more complicated when N$_2$O production is variable rather than when it is constant. N$_2$O production may vary with respect to production rate and/or $\delta$ values; in many aquatic environments, N$_2$O production is not likely to be constant. The N$_2$O production processes, nitrification and denitrification, are sensitive to redox conditions, which can be highly variable, due to diel changes in dissolved O$_2$ concentration, flow regime, etc. For example, [34] observed diel changes in the denitrification rate in the Iroquois River and Sugar Creek (Midwestern USA) and found that the denitrification was consistently greater during the day than night. The relative importance of nitrification and denitrification can change in response to the diel oxygen cycle: e.g., [46] observed a change from daytime nitrification to nighttime denitrification in a subtropical eutrophic stream. Coupling of N$_2$O and O$_2$ diel cycles has been observed in agricultural and waste-water treatment plant (WWTP) impacted rivers [36]. Since fractionation factors and substrates are different for nitrification and denitrification, ecosystem-scale fractionation factors may be rate and process dependent, and the $\delta$ values of N$_2$O production in a given ecosystem may not be constant over a diel cycle.

To simulate the diel variability, various scenarios were modelled by adjusting either production rate and/or the associated $\delta$ values. The variabilities in these input parameters were driven by a sine function with a 24 h period similar to a dissolve O$_2$ curve. In all scenarios, the chosen range of production rates was based on published N$_2$O flux rates (Table 1) and varied from 1 to...
5 mol/m$^2$/h$^1$ (Table 2), which was between the diel variation in N$_2$O flux observed by [33] and [46]. Temperature was held constant at 20°C. The value of $k$ was varied between 0.1 and 0.3 m/d (Table 2), within the range observed in other river studies (Table 1). The combination of production rates and $k$ values were chosen to produce N$_2$O between 150% and 500% saturation (Table 2) coinciding with the range of published data (Table 1). The range of $\delta$ values used for the N$_2$O source (Table 2) was within published values from various field studies [47]. For scenarios where the $\delta$ values of source N$_2$O was variable, the sine function for the $\delta$ values was synchronized so that maxima and minima $\delta^{15}$N values coincided with those of $\delta^{18}$O. This was done for simplicity, and because, in general, nitrification yields N$_2$O with lower $\delta^{15}$N and $\delta^{18}$O values than denitrification (e.g., [10,48]). Nevertheless, scenarios with greater amounts of N$_2$O reduction to N$_2$ can be modelled by increasing the source $\delta^{15}$N and $\delta^{18}$O values to those appropriate for any given ecosystem. Model scenarios were run until the output parameters (i.e., N$_2$O saturation and the $\delta$ values of dissolved, source, and emitted N$_2$O) reached dynamic steady state: model output was not constant over 24 h but the diel patterns on successive days were repeated.

Model Scenario #1: Variable Production Rate, Constant Isotopic Composition of Source

In scenario #1 (Table 2), the $\delta$ values of source N$_2$O were held constant and the production rate was variable. An example of such a system may be N$_2$O production via denitrification in river sediments with abundant NO$_3^-$. Denitrification rates in rivers have been observed to fluctuate in response to the diel O$_2$ cycle [34]. If the fractionation factors for denitrification are not rate dependent, the resulting N$_2$O production rate would be variable but the source $\delta$ values of N$_2$O values could be constant.

Here, the maximum concentration lagged approximately 2.75 h behind the maximum N$_2$O production rate, a function of the magnitude of the gas exchange coefficient, cf. [49]. The $\delta$ values for the instantaneously emitted N$_2$O were relatively constant and very similar to the N$_2$O source (within 0.4‰ for $\delta^{15}$N and 1.1‰ for $\delta^{18}$O, Figure 4, Table 3). However, the $\delta$ values of dissolved N$_2$O were more variable, spanning 16‰ for $\delta^{15}$N and 10‰ for $\delta^{18}$O. Thus, a change in the $\delta$ values of dissolved N$_2$O can be driven simply by a change in production rate and not necessarily a change in the $\delta$ values of the source. Since the system was at dynamic steady state, the average $\delta$ values of the emitted N$_2$O were identical to the average $\delta$ values of the source. This must be true in all steady-state cases to conserve the mass of the N$_2$O isotopologues.

In some aquatic systems, the N$_2$O production rate may remain constant with time but the $\delta$ values of the source may change with time. In rivers or lakes without a strong diel O$_2$ cycle, sediment denitrification may produce N$_2$O at an approximately constant rate. Denitrification rate may also be independent of water column NO$_3^-$ concentration if limited by factors other than diffusion in the sediments. The $\delta$ values of the source N$_2$O may thus change if the

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Table 1. Summary of relevant published data on N$_2$O production in aquatic environments.

| Location              | Range of N$_2$O Saturation (%) | Range of N$_2$O Flux (mol m$^{-2}$ h$^{-1}$) | Range of $k$ values (m h$^{-1}$) | Reference |
|-----------------------|--------------------------------|---------------------------------------------|---------------------------------|-----------|
| Ohio River, OH, US    | 95 to 745                      | 5 to 90                                    | —                               | [12]      |
| 5 agricultural streams, ON, CA (over 2–3 years) | 14 to 1700                     | 1 to 91.1                                  | 0.002 to 0.59                   | [5]       |
| 10 agricultural streams, ON, CA (over 17 diel cycles) | 30 to 2570                     | 0.33 to 52.1                               | 0.004 to 0.30                   | [45]      |
| Bang Nara River, TH   | 170 to 2000                    | —                                          | —                               | [20]      |
| LII River, NZ         | 201 to 404                     | 1.35 to 17.9                               | 0.13 to 0.82                    | [58]      |
| LII River, NZ         | 402 to 644                     | 0.46 to 0.89                               | 14.76                           | [33]      |
| Seine River, FR       | —                              | 2.2 to 5.2                                 | 0.04 to 0.06                    | [59]      |
| Canal Two, Yaqui Valley, MX | 100 to 6000                 | 0 to 34.9                                  | 0.3 to 0.6                      | [46]      |
| Agricultural stream, UK | 100 to 630                    | 0 to 37.5                                  | —                               | [60]      |
| Grand River, ON, CA   | 38 to 8573                     | 1.4 to 173.6                               | 0.06 to 0.35                    | [36,37]   |

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Table 2. Summary of input parameters for the SIDNO model scenarios for non-steady state production of N₂O.

| Scenario # | Results Figure | k (m/k) | N₂O Source Production Rate (µmol/m²/h) | N₂O Source δ¹⁵N and δ¹⁸O Values (%)
|------------|----------------|---------|----------------------------------------|---------------------------------------|
| Variable Production Rate, Constant Isotopic Composition of Source | 1 | 4 | 0.3 | 1 to 5 | −50, 10 |
| Constant Production Rate, Variable Isotopic Composition of Source | 2 | 5 | 0.3 | 3 | −50, 10 to −30, 10 |
| | 3 | 6 | 0.1 | 3 | −50, 10 to −30, 10 |
| Variable Production Rate, Variable Isotopic Composition of Source | 4* | 7 | 0.3 | 1 to 5 | −50, 10 to −30, 10 |
| | 5** | 8 | 0.3 | 1 to 5 | −50, 10 to −30, 10 |
| | 6* | 9 | 0.1 | 1 to 5 | −50, 10 to −30, 10 |

*Maximum production rate coincides with the lowest source δ values.
**Maximum production rate coincides with the highest source δ values.

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Figure 4. Model scenario #1 – isotopic composition of dissolved and emitted N₂O with a variable production rate and constant isotopic composition of the source. Note, in panel D, the data points for emitted N₂O are masked by the data point for source N₂O.
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Table 3. Summary of SIDNO output for model scenarios simulating non-steady state production of N₂O.

| Scenario # | Dissolved N₂O | Emitted N₂O |
|------------|---------------|-------------|
|            | Saturation    | \( \delta^{15}N, \delta^{18}O \) | \( \Delta \delta^{15}N \) | \( \Delta \delta^{18}O \) | \( \delta^{15}N \) | \( \delta^{18}O \) |
| Variable Production Rate, Constant Isotopic Composition of Source | | | | | | |
| 1          | 153 to 263    | -27.8, 24.6 to -12.1, 34.4 | 37.9 | 24.4 | -49.6, 9.5 to -50.2, 11.1 | 0.2 | 0.5 |
| Constant Production Rate, Variable Isotopic Composition of Source | | | | | | |
| 2          | 208           | -19.6, 29.4 to -3.7, to 37.4 | 30.4 | 19.4 | -45.3, 12.3 to -14.7, 27.7 | 4.7 | 2.3 |
| 3          | 423           | -26.1, 24.8 to -15.1, 30.3 | 23.9 | 14.8 | -37.2, 16.4 to -22.8, 23.6 | 12.8 | 6.4 |
| Variable Production Rate, Variable Isotopic Composition of Source | | | | | | |
| 4*         | 153 to 263    | -25.8, 25.6 to -1.2, 39.8 | 24.2 | 15.6 | -46.9, 11.3 to -18.0, 26.5 | 8 | 3.5 |
| 5**        | 153 to 263    | -11.2, 33.8 to -5.0, 36.1 | 38.8 | 23.8 | -41.6, 14.6 to -13.4, 28.1 | 8.4 | 4.6 |
| 6*         | 345 to 501    | -31.6, 21.6 to -18.4, 29.3 | 18.4 | 11.6 | -42.1, 13.7 to -29.4, 20.6 | 19.4 | 9.4 |

Temporally variable parameters are given as a range. \( \Delta \delta^{15}N \) and \( \Delta \delta^{18}O \) (\( \Delta = \delta_{\text{source}} - \delta_{\text{model}} \)) are the maximum difference between the range of source N₂O and the range for the model output parameter.

*Maximum production rate coincides with the lowest source \( \delta \) values.
**Maximum production rate coincides with the highest source \( \delta \) values.

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Figure 5. Model scenario #2 – Isotopic composition of dissolved and emitted with a constant production rate and variable isotopic composition of the source.
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\( \delta \) values of the NO\(_3^-\) substrate changed with time. For example, many studies have shown that the \( \delta \) values of residual NO\(_3^-\) increase during denitrification [50]. Similarly, NO\(_3^-\) from WWTPs may have different \( \delta \) values than agricultural runoff and diel changes in WWTP release may result in changing \( \delta \) values of NO\(_3^-\). Changes in N cycling may also vary on a diel basis but result in fortuitously similar N\(_2\)O production rates due to, for example, changes in the N\(_2\)O:N\(_2\) ratio of denitrification or changes in the relative importance of nitrification and denitrification. Thus changes in \( \delta \) values of the N\(_2\)O source do not necessarily indicate changes in N\(_2\)O production rates.

Model Scenario #2: Constant Production Rate, Variable Isotopic Composition of Source

In scenario #2, when the N\(_2\)O production rate was held constant and the \( \delta \) values of the source varied with time (from \(-50\%\) to \(-10\%\) for \( \delta^{15}\)N and from \(10\%\) to \(30\%\) for \( \delta^{18}\)O), the \( \delta \) values of the dissolved N\(_2\)O was also much farther from that of the source than the dissolved N\(_2\)O due to the effects of atmospheric exchange and the emitted N\(_2\)O varies linearly between the two source values. In contrast, the dissolved N\(_2\)O is parallel but offset from the line connecting the two sources (Figure 5, Table 3). The maximum difference between emitted and source N\(_2\)O was 4.7\% for \( \delta^{15}\)N and 2.3\% for \( \delta^{18}\)O. The dissolved and emitted \( \delta \) values also lagged 2.75 h behind the source as a result of gas exchange (as above). Since the system was at dynamic steady state, the average \( \delta \) values of the emitted N\(_2\)O were identical to the average \( \delta \) values of the source.

Model Scenario #3: Constant Production Rate, Variable Isotopic Composition of Source

To examine the effects of varying \( k \) on the scenario of constant N\(_2\)O production with variable isotopic signature of the source, \( k \) was reduced from 0.3 m/h (scenario #2) to 0.1 m/h (scenario #3; Figure 6, Table 3). The \( \delta \) values for the emitted N\(_2\)O were centred between the sources N\(_2\)O values, but dissolved N\(_2\)O \( \delta \) values were farther from tropospheric N\(_2\)O than the high-\( k \) scenario #2 (Figure 6 D).

The effect of reducing \( k \) was an increase in N\(_2\)O concentration with the same production rate and a shift in the \( \delta \) values of dissolved N\(_2\)O toward the source values. Reducing \( k \) also dampened the response between the instantaneous \( \delta \) values of the emitted N\(_2\)O and the \( \delta \) values of the source. As above, the lag time between the \( \delta \) values of the source and emitted N\(_2\)O increased as \( k \) decreased. The total range of the source and
emitted \( \delta \) values decreased. The difference between the source and emitted \( \delta \) values was 12.8\% for \( \delta^{15}N \) and 6.4\% for \( \delta^{18}O \).

To simulate a system alternating between two \( \text{N}_2\text{O} \) production processes, such as differing relative contributions of nitrification and denitrification, with different rates of \( \text{N}_2\text{O} \) production and \( \delta \) values, the model was run with both production rate and its \( \delta \) values variable with time (scenarios #4, #5, and #6). The production rate and \( \delta \) values were adjusted so that the maximum rate coincided with the lowest source \( \delta \) values in scenarios #4 and #6 and so that maximum rate coincided with the highest source \( \delta \) values in scenario #5.

**Model Scenario #4: Variable Production Rate, Variable Isotopic Composition of Source**

For scenario #4, the resulting \( \text{N}_2\text{O} \) concentrations were identical to those in model scenario #1, with the maximum concentration lagging approximately 2.75 h behind the maximum production rate (Figure 7, Table 3). The relationship between the \( \delta \) values of the dissolved and emitted \( \text{N}_2\text{O} \) was more complex than in other scenarios. The lag time between the maximum source \( \delta \) values and those of dissolved and emitted \( \text{N}_2\text{O} \) (when the production rate was minimum) was 3.75 h; however, the lag time between the minimum source \( \delta \) values and those of the dissolved and emitted \( \text{N}_2\text{O} \) (when the production rate was maximum) was only 2.25 h. The difference between the emitted and source \( \text{N}_2\text{O} \) was 3.1\% to 8.0\% for \( \delta^{15}N \) and 1.3\% to 3.4\% for \( \delta^{18}O \). The \( \delta \) values of emitted \( \text{N}_2\text{O} \) were closer to those of the source during periods of high production rates (and thus higher concentrations) than periods of low production rates. However, the flux-weighted average \( \delta \) values of emitted \( \text{N}_2\text{O} \) were equal to the average production-weighted source \( \delta \) values because the system was at dynamic steady state.

**Model Scenario #5: Variable Production Rate, Variable Isotopic Composition of Source**

The isotopic counterpoint to scenario #4 is adjusting the timing of maximum \( \text{N}_2\text{O} \) production to coincide with the highest \( \delta \) values of production (scenario #5). All other parameters were the same as scenario #4 (Table 3). The resulting pattern for the \( \delta \) values of dissolved \( \text{N}_2\text{O} \) was very different than scenario #4 (Figure 8, Table 3). While the dissolved \( \text{N}_2\text{O} \) concentrations were identical to the model scenario #4, the \( \delta \) values of dissolved \( \text{N}_2\text{O} \) were nearly constant with time. The relationship between the \( \delta \) values of emitted and source \( \text{N}_2\text{O} \) was similar to scenario #4, although the instantaneous difference in \( \delta \) values were slightly greater. The \( \delta \) values of the dissolved \( \text{N}_2\text{O} \) were greatly dampened by the fact...
that maximum production rate coincided with source \( \delta \) values that were closest to tropospheric N\(_2\)O. In scenario #4, the high rates of N\(_2\)O production at \( \delta \) values very different than tropospheric N\(_2\)O increased the amplitude of the \( \delta \) values of dissolved N\(_2\)O.

**Model Scenario #6: Variable Production Rate, Variable Isotopic Composition of Source**

To determine the effects of a lower \( k \) on model scenario #4, \( k \) was reduced from 0.3 m/h to 0.1 m/h for scenario #6. As shown above, lower \( k \) increased the dissolved N\(_2\)O concentrations and dampened the diel range of \( \delta \) values of both dissolved and emitted N\(_2\)O (Table 3, Figure 9). Lower \( k \) also increased the lag time between the \( \delta \) values of emitted and source N\(_2\)O and increased the difference between the \( \delta \) values of emitted and source N\(_2\)O (Figure 9). As in all scenarios, the flux-weighted average \( \delta \) values of emitted N\(_2\)O were equal to the average production-weighted source \( \delta \) values.

**Grand River**

The ability of SIDNO to reproduce measured patterns of N\(_2\)O concentration and \( \delta \) values in a human-impacted river was also assessed. The Grand River is a seventh-order, 300 km long river that drains 6800 km\(^2\) in southern Ontario, Canada, into Lake Erie, see [36,37,51]. There are 30 WWTPs in the catchment and their cumulative impact can be observed via the increase in artificial sweeteners in the river [52].

Samples were collected approximately hourly for 28 h at two sites in the central, urbanized portion of the river: sites 9 and 11 in [51,52]. The upstream site, Bridgeport, is where the river enters the urban section of the river at the city of Waterloo and is immediately above that city’s WWTP. Blair is 26.6 km downstream of Bridgeport and below the cities of Waterloo and Kitchener. It is also 5.5 km downstream of the Kitchener WWTP. Average river depth at both sites was 30 cm. Values of \( k \) were determined by best-fit modelling of diel O\(_2\) and \( \delta^{18}\)O-O\(_2\) values at the sites [36]. N\(_2\)O concentration analyses were performed on a Varian CP-3800 gas chromatograph with an electron capture detector and isotopic ratio analyses were performed on a GV TraceGas pre-concentrator coupled to a GV Isoprime isotope ratio mass spectrometer, see [5] for analytical details.

Data from upstream and downstream of large urban wastewater treatment plants on the Grand River show diel patterns in N\(_2\)O saturation and \( \delta \) values (Figures 10 and 11). At the Bridgeport site, the diel patterns of N\(_2\)O saturation and \( \delta^{15}\)N values were opposite of each other, that is, when N\(_2\)O saturation was highest around sunrise the \( \delta^{15}\)N values were lowest and when

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**Figure 8. Model scenario #5 – Isotopic composition of dissolved and emitted N\(_2\)O with a variable production rate and variable isotopic composition of the source.** Maximum production rate is in sync with the greatest \( \delta^{15}\)N and \( \delta^{18}\)O values of the source. doi:10.1371/journal.pone.0090641.g008
when $N_2O$ saturation was lowest around before sunset the $\delta^{15}N$ values were greatest. $\textit{R}^2$ values between field and model data for $N_2O$ saturation, $\delta^{15}N$, and $\delta^{18}O$ values are 0.83, 0.68, and 0.30. Model results reproduce the range and sinusoidal patterns of the field data though the $\delta^{18}O$ fit was poor in the second half of the field data. The diel pattern in $\delta^{18}O$ values was similar to that of $\delta^{15}N$ but was shifted earlier by about 4 h. These patterns were similar to those of scenario #4 (variable $N_2O$ production and variable $\delta$ values of the source $N_2O$ coinciding when maximum production rates coincided with lowest source $\delta$ values) and the result of consistent diel five-fold variability in $N_2O$ production and variability in the $\delta^{18}O$ of the $N_2O$ produced in the river.

At the downstream Blair site, both $\delta^{15}N$ and $\delta^{18}O$ values were much lower and exhibited a greater range than at Bridgeport. $\textit{R}^2$ values between field and model data for $N_2O$ saturation, $\delta^{15}N$, and $\delta^{18}O$ values are 0.78, 0.53, and 0.03. Model results reproduce the range and peak-and-trough pattern of the $N_2O$ saturation and $\delta^{15}N$ data. Model results reproduce the range of $\delta^{18}O$ values but the pattern is not well reproduced. While all data at Bridgeport exhibited smooth, sinusoidal diel changes, the data at Blair show rapid changes. The diel patterns of $N_2O$ saturation and $\delta^{15}N$ values were opposite of each other, that is, when $N_2O$ saturation was highest around midnight, the $\delta^{15}N$ values were lowest and when when $N_2O$ saturation was lowest during mid-day, the $\delta^{15}N$ values were greatest. The diel pattern in $\delta^{18}O$ values was more complex at Blair than at Bridgeport suggesting that daytime and nighttime were associated with different $\delta^{18}O$ values of $N_2O$ production. These patterns were similar to those of scenario #5 (variable $N_2O$ production and variable $\delta$ values of the source $N_2O$ coinciding when maximum production rates coincided with highest source $\delta$ values) and the result of a five-fold variability in day-to-night $N_2O$ production and variability in $\delta^{15}N$ and $\delta^{18}O$ of the $N_2O$ produced in the river.

At the downstream Blair site, both $\delta^{15}N$ and $\delta^{18}O$ values were much lower and exhibited a greater range than at Bridgeport. $\textit{R}^2$ values between field and model data for $N_2O$ saturation, $\delta^{15}N$, and $\delta^{18}O$ values are 0.83, 0.68, and 0.30. Model results reproduce the range and sinusoidal patterns of the field data though the $\delta^{18}O$ fit was poor in the second half of the field data. The diel pattern in $\delta^{18}O$ values was similar to that of $\delta^{15}N$ but was shifted earlier by about 4 h. These patterns were similar to those of scenario #4 (variable $N_2O$ production and variable $\delta$ values of the source $N_2O$ coinciding when maximum production rates coincided with lowest source $\delta$ values) and the result of consistent diel five-fold variability in $N_2O$ production and variability in the $\delta^{18}O$ of the $N_2O$ produced in the river.

For both Bridgeport and Blair data, the cause of poorer fits for $\delta^{18}O$ than $\delta^{15}N$ deserve further research. Adding concomitant measurements of $\delta^{15}N$ and $\delta^{18}O$ values of $NO_3^-$ may provide clues about N cycling and help explain some of the observed variability in $N_2O$ [53]. Predicting $\delta^{18}O$-$N_2O$ values from nitrification [11] and denitrification [54] is difficult because of the complex relationship between $\delta^{18}O$-$H_2O$ values and $\delta^{18}O$-$N_2O$ values. Additionally, diel variability in $N_2O$ reduction to $N_2$ [45,46], may also manifest itself in $\delta^{18}O$-$N_2O$ values because of the strong O isotope fractionation factor during denitrification [55].
Discussion

Calculating the δ values of emitted or source N₂O is critical for regional and global N₂O isotopic budgets and also provides information about the source of N₂O and thus N cycling processes. However, SIDNO can simulate the relationships between the δ values of dissolved, source, and N₂O emitted from aquatic ecosystems to the atmosphere. In systems with N₂O production at dynamic steady state, the δ values of dissolved N₂O will not always be directly indicative of the δ values of the source N₂O. The difference between dissolved and source δ values increases as N₂O saturation decreases (as demonstrated in Figures 2 and 3). Even above 1000% saturation (from high production rates and/or low k), the δ values of dissolved N₂O will only approach δ values of the source but offset by 0.7% for δ¹⁵N and 1.9% for δ¹⁸O, a result of the ev values (Figure 3). At constant N₂O production rates and δ values, the source and emitted δ values can be quantified since the δ values of emitted N₂O must be identical to those of the source and can be calculated from dissolved values (Figure 3; equations 5 and 6).

Our modelling results identified the limitations associated with simple interpretation of dissolved N₂O isotope data since the δ values of dissolved and emitted data are synchronous but rarely offset by a constant value. If N₂O saturation changes with time, the N₂O production rate must also have changed with time, provided k had been constant (compare model scenarios #1 and #2 in Figures 4 and 5). In contrast, changes in δ values of dissolved N₂O do not require a change in the source δ values (model scenario #1 and Figure 4), while constant δ values of dissolved N₂O do not require constant source δ values (for example model scenario #5 in Figure 8).

When δ values of the source N₂O are variable, the relationship between emitted and source N₂O becomes complicated. The δ values of emitted N₂O will lag behind those of the source and the amplitude of the diel range of δ values will be dampened relative to the source. The amount of lag and dampening is a function of k, N₂O production rate and timing, and the proximity of the source δ values to those of the atmosphere (compare Figures 2 with 3 and Figures 4 with 6). Qualitatively, the δ values of emitted N₂O will be similar to the source if the equilibration time of dissolved N₂O is small relative to the period of source variability (e.g., 24 h period due to diel changes in N cycling [36,45]). Assuming homogeneous N₂O release upstream, the equilibration time can be approximated from a decay curve as \( \frac{3z}{k} \), where z is mean depth [49]. If z is small and/or k is high, the equilibration time will be short and the δ values of the emitted N₂O will be close to the source. With decreasing k (or increasing equilibration time), the δ values of emitted N₂O will lag farther behind and will always have a smaller range of δ values than the source. At the most extreme case, the variability in the δ values of emitted N₂O will be reduced to nearly

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**Figure 10.** Diel variability in N₂O concentration and δ values at Bridgeport in the Grand River, Canada. The time axis begins at 00:00 on 2007-06-26. Maximum production rate is in sync with the greatest δ¹⁸O values of the source, while δ¹⁵N of the source was constant. R² values between field and model data for N₂O saturation, δ¹⁵N, and δ¹⁸O values are 0.83, 0.68, and 0.30. This is similar to model scenario #4 (Figure 7). doi:10.1371/journal.pone.0090641.g010

**Figure 11.** Diel variability in N₂O concentration and δ values at Blair in the Grand River, Canada. The time axis begins at 00:00 on 2007-06-26. Maximum production rate is in sync with the lowest δ¹⁵N and δ¹⁸O values of the source. R² values between field and model data for N₂O saturation, δ¹⁵N, and δ¹⁸O values are 0.78, 0.53, and 0.03. This is similar to model scenario #5 (Figure 8). doi:10.1371/journal.pone.0090641.g011
Dissolved N₂O Isotopes in Aquatic Ecosystems

Table 4. Summary of the results of the SIDNO modelling as the predictive relationship between observations and implications.

| Observed Parameter                                                                 | Implications                                                                 | Examples                                                                 |
|-----------------------------------------------------------------------------------|------------------------------------------------------------------------------|-------------------------------------------------------------------------|
| Dissolved N₂O concentration is constant with time                                  | The N₂O production rate is constant with time (if k and temperature are also constant). The N₂O flux to the atmosphere is equal to the production rate. This may not be true if the concentration is close to atmospheric equilibrium. | Scenarios #2 and #3                                                      |
| Dissolved N₂O concentration is variable with time in a sinusoidal pattern          | The N₂O production rate is variable with time (if concentration change cannot be explained by change in k or temperature). The average N₂O flux to the atmosphere is equal to the average production rate. | Scenario #1                                                              |
| δ¹⁵N and δ¹⁸O of dissolved N₂O is constant with time                               | The observation is inconclusive. At concentrations near atmospheric equilibrium, isotopic composition of dissolved N₂O will approximate tropospheric N₂O, regardless of source values. A constant isotopic signature of dissolved N₂O that is different from tropospheric N₂O can indicate either a constant source (if production rate is constant), or a variable source. | Scenario #5                                                              |
| Calculated δ¹⁵N and δ¹⁸O of emitted N₂O is constant with time                     | The isotopic composition of the source is constant with time, and equal to the calculated value for emitted N₂O | Scenario #1                                                              |
| Calculated δ¹⁵N and δ¹⁸O of emitted N₂O is variable with time                    | The range in δ¹⁵N and δ¹⁸O of emitted N₂O is the minimum for the range in that of the source. The flux weighted average δ¹⁵N and δ¹⁸O of emitted N₂O is equal to the production weighted average source values. | Scenarios #2–#6                                                          |
| Long residence time relative to variability of source (need to independently determine k) | The changes in N₂O concentration and δ¹⁵N and δ¹⁸O of emitted N₂O will be dampened relative to, and lag behind, that of the source | Scenarios #3 and #6                                                      |
| Short residence time relative to variability of source (need to independently determine k) | The changes in N₂O concentration and δ¹⁵N and δ¹⁸O of emitted N₂O will be indicative of changes in the source | Scenarios #1, #2, #3 and #4                                               |

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zero and δ values of the emitted N₂O would be equal to the average production-weighted source δ values. At very long equilibration times, the probability of N₂O consumption increases, a process not explicitly included in SIDNO where the δ value of the source N₂O is simply that which is released to the water column.

Separating N₂O production into nitrification and denitrification requires independent knowledge about the δ values of the source N and O in aquatic ecosystems. It is therefore not possible to state a single δ¹⁵N value for nitrification–N₂O and one for denitrification–N₂O applicable to all aquatic ecosystems. The δ¹⁵N value of the N₂O precursors NH₄⁺ and NO₃⁻ vary across ecosystem as a result of human impact and N loading (agricultural and WWTP) as well as the source of N, and additional N transformations in the aquatic ecosystem. For example, along the length of the Grand River, δ¹⁵N values of NH₄⁺ and NO₃⁻ exhibit systematic trends resulting from the confluence of agricultural tributaries and large urban waste-water treatment plants (Schiff et al., unpublished results, [53]). Nevertheless, these values can be measured and biogeochemical relationships between N species, redox, and N₂O can be used as supporting information for process separation (e.g., [5,12,36,43]). The δ¹⁵O value of N₂O will also vary across ecosystems as a result of its close relationship with δ¹⁸O–H₂O and to a lesser extent δ¹⁸O–O₂ [10,11,54,56]. Fortunately, δ¹⁸O–H₂O values can be easily predicted and measured [57]. Thus, once δ values of N₂O precursors have been identified, biogeochemical data can provide an indication about the diel pattern of N₂O production processes, and ranges of potential end-member δ values can be calculated (e.g., [5] summarize isotopic fractionation for δ¹⁵N and [10,11,54,56] for δ¹⁸O) and the model used to fit the field data.

Conclusions

In aquatic ecosystems, the instantaneous δ values of N₂O emitted to the atmosphere are easily calculated if the water temperature and dissolved N₂O concentration and δ values are known. Our modelling efforts illustrate that complex relationships exist between dissolved and source N₂O and that the δ values of dissolved N₂O are not always representative of either the N₂O produced or emitted to the atmosphere. Thus, calculated δ values of the emitted N₂O are the values that should be used to draw conclusions about N₂O emission from aquatic systems and the global N₂O cycle rather than the more commonly used instantaneous values (Table 4). The flux-weighted δ values of emitted N₂O can provide average production-weighted δ values of the N₂O source under dynamic steady-state in aquatic ecosystems.

If the δ values of emitted N₂O are constant with time, either the δ values of the source must also be constant or the N₂O equilibration time is very long. However, if the calculated δ values of emitted N₂O vary with time then the δ values of the source must also vary with time producing a diagnostic pattern. These findings are more robust than using dissolved δ values alone since dissolved
δ values can change simply with a change in N₂O production rate, changes in source δ values, and changes in k, N₂O residence time, dependent on production rate, k, and z, will determine the lag time between the δ values of emitted and source N₂O. The difference in timing between maxima and minima δ values of emitted N₂O and the maxima and minima of dissolved N₂O is indicative of how the δ values of the source change. For all these reasons, we urge caution when using single samples of N₂O concentration and δ to calculate fluxes of N₂O to the atmosphere and inferring N₂O production pathways.

Ultimately, the dynamic model SIDNO may be used to estimate k, N₂O production rate and δ values of the N₂O source, an indication of the production pathway and N cycling, in aquatic ecosystems via inverse modelling. If physical properties, such as depth and temperature are known, SIDNO may be used to fit the measured field data (N₂O concentration and δ values) by adjusting the N₂O source parameters. SIDNO can also be used to explore the dynamics between dissolved, source, and emitted N₂O to query production scenarios and design field campaigns for studies of N cycling processes.

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Author Contributions

Conceived and designed the experiments: SJT JJV SLS. Performed the experiments: SJT JJV. Analyzed the data: SJT JJV SLS. Contributed reagents/materials/analysis tools: SJT JJV SLS. Wrote the paper: JJV SJT SLS.

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