Photosynthetic responses of *Halimeda scabra* (Chlorophyta, Bryopsidales) to interactive effects of temperature, pH, and nutrients and its carbon pathways

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In this study, we evaluated the interactive effects of temperature, pH, and nutrients on photosynthetic performance in the calcareous tropical macroalga *Halimeda scabra*. A significant interaction among these factors on gross photosynthesis (\( P_{\text{gross}} \)) was found. The highest values of \( P_{\text{gross}} \) were reached at the highest temperature, pH, and nutrient enrichment tested and similarly in the control treatment (no added nutrients) at 33°C at the lowest pH. The \( Q_{10} P_{\text{gross}} \) values confirmed the effect of temperature only under nutrient enrichment scenarios. Besides the above, bicarbonate (\( \text{HCO}_3^- \)) absorption was assessed by the content of carbon stable isotope (\( \delta^{13}C \)) in algae tissue and by its incorporation into photosynthetic products, as well as by carbonic anhydrase (CA) inhibitors (Acetazolamide, AZ and Ethoxyzolamide, EZ) assays. The results of \( \delta^{13}C \) revealed this species uses both, \( \text{CO}_2 \) and \( \text{HCO}_3^- \) forms of \( C_i \) relying on a \( \text{CO}_2 \) Concentration Mechanism (CCM). These results were validated by the EZ-AZ inhibition assays in which photosynthesis inhibition was observed, indicating the action of internal CA, whereas AZ inhibitor did not affect maximum photosynthesis (\( P_{\text{max}} \)). The incorporation of \( ^{13}C \) isotope into aspartate in light and dark treatments also confirmed photosynthetic and non-photosynthetic the \( \text{HCO}_3^- \) uptake.
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ABSTRACT

In this study, we evaluated the interactive effects of temperature, pH, and nutrients on photosynthetic performance in the calcareous tropical macroalga *Halimeda scabra*. A significant interaction among these factors on gross photosynthesis ($P_{\text{gross}}$) was found. The highest values of $P_{\text{gross}}$ were reached at the highest temperature, pH, and nutrient enrichment tested and similarly in the control treatment (no added nutrients) at 33°C at the lowest pH. The $Q_{10}$ $P_{\text{gross}}$ values confirmed the effect of temperature only under nutrient enrichment scenarios. Besides the above, bicarbonate (HCO$_3^-$) absorption was assessed by the content of carbon stable isotope ($\delta^{13}$C) in algae tissue and by its incorporation into photosynthetic products, as well as by carbonic anhydrase (CA) inhibitors (Acetazolamide, AZ and Ethoxyzolamide, EZ) assays. The results of $\delta^{13}$C revealed this species uses both, CO$_2$ and HCO$_3^-$ forms of $C_i$ relying on a CO$_2$ Concentration Mechanism (CCM). These results were validated by the EZ-AZ inhibition assays in which photosynthesis inhibition was observed, indicating the action of internal CA, whereas AZ inhibitor did not affect maximum photosynthesis ($P_{\text{max}}$). The incorporation of $^{13}$C isotope into aspartate in light and dark treatments also confirmed photosynthetic and non-photosynthetic the HCO$_3^-$ uptake.

Key index words: Carbonic anhydrase, CCM, $^{13}$C isotope, $\delta^{13}$C, *Halimeda scabra*, interactive effects, nutrients, pH, photosynthesis, $Q_{10}$, temperature.

INTRODUCTION

Photosynthetic parameters respond faster to environmental changes than algae C and N content, hence their usefulness in short-term studies (Figueroa et al., 2009). Photochemical and biochemical reactions of photosynthesis continually respond to environmental conditions. Irradiance, temperature, and nutrient concentration including CO$_2$ levels are among the main environmental factors limiting photosynthesis (Raven & Hurd, 2012; Zweng, Koch & Bowes, 2018). Algal ecophysiology studies have traditionally quantified temperature dependence using the metabolic quotient $Q_{10}$, which describes the metabolic increase accompanied by an increase of 10°C in an optimal temperature range (Bruno, Carr & O’Connor, 2015; Vásquez-Elizado & Enríquez, 2016). This quotient $Q_{10}$ has also been used as a proxy to analyze the effect of
temperature on nutrient absorption where it was found that by doubling the temperature the rate of nutrient absorption is doubled (Harrison & Hurd, 2001).

For aquatic plants, another limiting factor for photosynthesis is CO₂, since it is the only source of carbon that can be assimilated by the Ribulose 1,5 bisphosphate carboxylase oxygenase enzyme (RuBisCO) (Falkowski & Raven, 2007). At seawater pH (8.1 – 8.3) CO₂ is only between 0.5 – 1% of all dissolved inorganic carbon, while more than 91% is in the form of HCO₃⁻ and the remaining 8% is in the form of CO₃²⁻ (Hurd et al., 2009; Diaz-Pulido et al., 2016). Moreover, since the diffusion of CO₂ through the cell membrane is slower in water than in air; many algae and higher plants have acquired mechanisms that promote intracellular CO₂ accumulation, allowing photosynthetic organisms to reduce carbon limitation by increasing the concentration of CO₂ in the vicinity of RuBisCO (CO₂ Concentration Mechanisms, CCM). Parallel to this, CCM’s contribute to decreasing photorespiration due to the oxygenase activity of RuBisCO (Ogren, 1984; Enríquez & Rodríguez-Román, 2006; Cornwall, Revill & Hurd, 2015). In general, most algae can acquire inorganic carbon (Cᵢ) for RuBisCO through diffusion and active absorption of both, CO₂ and HCO₃⁻ (Badger & Price, 1994; Giordano, Beardall & Raven, 2005; Hurd et al., 2009). In many cases, the activity of CCMs has been associated with the direct or indirect use of HCO₃⁻ (Reiskind, Seamon & Bowes, 1988; Invers et al., 2001; Enríquez & Rodríguez-Román, 2006). Some macroalgae convert bicarbonate (HCO₃⁻) into CO₂ extracellularly with carbonic anhydrase (CA) thus CO₂ enters the cell by active transport or diffusion. Other algae incorporate HCO₃⁻ actively through the cell membrane and, intracellularly, an internal CA converts HCO₃⁻ into CO₂ (Badger & Price, 1994). The activity of carbonic anhydrases has been widely documented in algae (Reiskind, Seamon & Bowes, 1988; Invers, Perez & Romero, 1999; Enríquez & Rodríguez-Román, 2006) and they play a significant role in CCM’s.

Many studies have examined the combined effects of environmental variables on algae photosynthetic responses: CO₂ and temperature (Campbell et al., 2016; Kram et al., 2016; Vásquez-Elizondo & Enríquez, 2016); CO₂ and light (Vogel et al., 2015); light and nutrients (Zubia, Freile-Pelegrín & Robledo, 2014); CO₂ and nutrients (Hofmann et al., 2014; Hofmann et al., 2015; Bender-Champ, Diaz-Pulido & Dove, 2017); CO₂, nutrients, and temperature (Stengel et al., 2014), and CO₂, nutrients and light (Celis-Plá et al., 2015). Multiple stressors could have an interactive influence causing complex responses at the physiological and ecological level (Hofmann et al., 2014), which makes them difficult to interpret. Therefore, studies that combine ocean acidification scenarios with other factors such as temperature, light, and nutrient availability are particularly necessary since changes in these parameters are co-occurring with changes in carbonate chemistry in the seawater (Harley et al., 2012; Hofmann et al., 2014).

Halimeda is a calcifying genus of siphonous green algae (Bryopsidales, Chlorophyta) which are important components of tropical and subtropical reefs and lagoons. Some species of this genus often appear dominating Caribbean coral reefs (Beach et al., 2003; Hofmann et al., 2014) where they contribute as primary producers, food source and habitat, sand production, and coral-reef formation. Halimeda scabra Howe is particularly abundant in the front reef and
shallow rocky areas of the Caribbean Reefs (Alcolado et al., 2003). Despite the ecological studies above-mentioned, to our knowledge, no previous physiological studies have been reported for this species.

Photosynthetic responses to the combined effect of environmental variables have been studied in some Halimeda species, for example in *H. opuntia*, the effect of nutrients and pH (Hofmann et al., 2014; Hofmann et al., 2015), in *H. incrassata* and *H. simulans* the effect of pH and temperature (Campbell et al., 2016), and in *H. opuntia* the effect of pH and light (Vogel et al., 2015). These studies have suggested that an increase in both, CO$_2$ (low pH) and high temperature could have a positive synergistic effect on photosynthetic rates (Kram et al., 2016). However, *Halimeda* responses to high CO$_2$ have been diverse; in some species a decrease in photosynthesis with the reduction of pH has been observed (Price et al., 2011; Sinutok et al., 2012; Meyer et al., 2016) others have shown the opposite effect (Peach, Koch & Blackwelder, 2016) or a lack of a significant response (Price et al., 2011; Campbell et al., 2016). In general, there are still insufficient studies on the physiology of the genus *Halimeda* that allow us to understand the diversity of physiological responses to the interactive effects of environmental variables and the mechanisms involved in those responses.

In this study, we hypothesized that a synergistic increase in environmental factors (temperature, pH, and nutrients) would enhance *H. scabra* photosynthesis, which absorbs bicarbonate supported by a CCM. We evaluated the interactive effect of temperature, pH, and nutrient levels on photosynthetic responses of *H. scabra*. Additionally, we determined the $C_i$ uptake mechanisms by measuring the effect of CA inhibitors on $P_{\text{max}}$, analyzing $\delta^{13}$C values, and evaluating the incorporation of stable isotope $^{13}$C into resulting products of photosynthesis.

**MATERIALS AND METHODS**

**Biological material and culture conditions**

*H. scabra* was collected in February 2017 in Xcalacoco, Quintana Roo, Mexico (20.660035 N, -87.034655 W), where it grows over rocky substrates between 1.5 and 2.0 m depth. In this area there are two marked seasons: a dry season from November to May with mean seawater temperatures of ~24°C and a rainy season from June to October, with mean ~30°C reaching extreme values of 33°C, coinciding with summer, while the mean annual seawater temperature is ~28°C (Robledo & Freile-Pelegrín, 2005; Álvarez-Cadena et al., 2007; Rodríguez-Martínez et al., 2010). The area is also characterized by submarine groundwater discharges to the coastal environment a pathway for nutrients transport from land to the marine environment (NO$_3^-$, NO$_2^-$, NH$_4^+$, SRP, SRSi) conferring a pulsatile performance depending on the season, which can be particularly high in some localities especially during rainy season with pH important variations ranging from 7.0 to 8.5 (Crook et al., 2012; Hernández-Terrones et al., 2015).

Taxonomic determination of specimens was done according to Hillis-Colinvaux (1980). The algae were transported to the laboratory in cool boxes. At the laboratory, samples were cleaned with seawater to remove epiphytes and placed in 12 L aquarium with filtered seawater (36 PSU, pH 8.2) and kept under constant aeration at 24°C. Irradiance was set at 115 μmol photons m$^{-2}$ s$^{-1}$ provided by fluorescent lamps under a 12:12 hours light-dark photoperiod.
**Photosynthetic measurements**

To test the effect of temperature, pH, and nutrient levels on *H. scabra* photosynthesis, a three-factorial design with 36 combinations was used (Zar, 1996) (Table S1). The following treatments and levels were tested: (1) temperature at three levels (24, 28 and 33ºC) maintained constant by placing the BOD bottles in a water bath connected to a temperature controlled water recirculation system (Cole-Parmer® Polystat® Refrigerated Recirculator, USA); (2) pH at three levels (7.5, 8.2, and 8.6) obtained by the addition of 0.5 M HCl or 0.5 M NaOH solutions (Lignell & Pedersén, 1989; Invers et al., 2001; Zou, 2014); (3) nutrient concentrations (KNO$_3$:K$_2$PO$_4$) evaluated at four levels: low (1:0.1 µM), medium (5:0.5 µM), high (10:1.0 µM) and (4) control treatment without nutrients added to seawater. These temperature, pH and nutrient levels were selected according to prevailing conditions at the collecting site, as described above. To prepare the different combinations first seawater pH was adjusted, and 48 hours later the pH was measured again and readjusted when necessary, after that the nutrients were added according to the required nutrient concentration (Lignell & Pedersén, 1989).

Photosynthetic responses were evaluated by the light-dark bottles method following the oxygen evolution versus irradiance (P-E curves) according to Thomas (1988) using a YSI 5000 Dissolved Oxygen Meter with YSI 5905 BOD Probe (YSI Incorporated Yellow Springs, Ohio, USA). To minimize wound effects, thalli were cut off and weighed 24 hours before oxygen determinations. Apical fragments (0.1 g wet weight) were placed in 60 ml Biological Oxygen Demand (BOD) bottles containing the seawater previously prepared according to each of the 36 combinations. For each treatment seven bottles ($n = 7$) were used plus one blank, bottle filled with seawater only. Each combination was assessed on separated days with different fragments of algae tissues and according to each temperature level established in Table S1.

Once the temperature was set to the corresponding treatment the algae were exposed during one hour to each of seven successive irradiances selected (0, 100, 170, 200, 272, 436, 770 µmol photons m$^{-2}$ s$^{-1}$) generated by a 500 W halogen lamp and using different mesh size filters until darkness. Irradiances were measured with a spherical underwater quantum sensor (LI-193SA) connected to LI-1500 Light Sensor Logger (LI-COR, Nebraska, USA). The maximum photosynthesis rate ($P_{max}$) was calculated as the average of the three highest oxygen production values at saturation irradiances. The dark respiration rate ($R_d$) was determined as oxygen consumption in total darkness, while the gross photosynthesis ($P_{gross}$) was determined as net photosynthesis plus dark respiration. At the end of each assessment the dry weight (DW) was determined, whereby the results were expressed as mg of oxygen g dry weight h$^{-1}$ in 300 ml. All determinations were performed using Instant Ocean® synthetic seawater (Marineland, USA), prepared using distilled water and sterilized by autoclaving, this water was free of nitrate and phosphate therefore nutrient were added accordingly to the levels of each experimental treatments (Table S1).

**Effect of temperature on $P_{gross}$: photosynthetic $Q_{10}$ coefficient**

To better understand photosynthetic responses to temperature we calculated the photosynthetic quotient $Q_{10}$ of $P_{gross}$ under different pH and nutrient conditions. The
photosynthetic quotient was determined as the change in the photosynthetic rate within a rise in temperature of 5°C, from 28°C (T1) to 33°C (T2) according to the following formula:

\[ Q_{10} = \left( \frac{\text{Rate}_2}{\text{Rate}_1} \right)^{\frac{10}{T_2 - T_1}} \]

where, Rate 1 and Rate 2 were reaction rates measured at temperatures T1 and T2, respectively (Wernberg et al., 2016).

Inorganic carbon pathways

Bicarbonate (HCO$_3^-$) uptake for photosynthesis were assessed through three techniques: (1) carbon stable isotope ($\delta^{13}$C) values in algal tissue (2) CA inhibitor effects on $P_{\text{max}}$, and (3) $^{13}$C stable isotope uptake and its incorporation into resulting products of photosynthesis.

Carbon stable isotope ($\delta^{13}$C) values in tissue from field samples

Whole thalli were carefully washed and decalcified in hydrochloric acid (HCl) at 0.6 M for 8 hours, with hourly changes until full bubbling cessation. Afterward, the material was rinsed with distilled water and dried for 24 hours at 70°C. The dried material was ground in a mortar and sieved. Samples of five mg were weighed on analytical balance (precision of 0.0001 g) and individually packaged in microcapsules (5 x 9 mm) for mass spectrophotometer isotopic analysis in the Stable Isotropy laboratory at the University of California at Davis, CA, USA.

Carbonic anhydrase inhibition assays

Two CA inhibitors were used in this study: a) dextran-bound acetazolamide (AZ) that does not penetrate into the cell and only inhibits extracellular CA (Bjork et al., 1992), and b) 6-ethoxyzolamide (EZ) that penetrates through the cell wall and membranes, and inhibits both external and internal CA (Bjork et al., 1992). AZ and EZ were dissolved in 0.05 N NaOH to a final concentration of 0.1 g ml$^{-1}$ and 10 mM respectively (Bjork et al., 1992). Experimental treatments were prepared with seawater from the collecting area filtered with vacuum pump and sterilized by autoclaving. The inhibitors were added to the experimental seawater before the incubations to obtain a final inhibitor concentration of 100 µM (Bjork et al., 1992). Photosynthesis rates were tested under four treatments: 1) addition of AZ; 2) the addition of EZ; 3) the combination of both, AZ and EZ and 4) a control treatment with seawater without inhibitors. Maximum photosynthesis (P-E curves) was measured as previously described but at 28°C of temperature ($n = 7$).

$^{13}$C Labeling for the incorporation of NaH$^{13}$CO$_3$ into photosynthetic products

Initially, inorganic carbon was removed from filtered and sterilized seawater by reducing pH to $\sim 4$ adding HCl 0.5 M and nitrogen bubbling for 5 hours, subsequently the pH was raised to 8.2 adding NaOH 0.5 M (Invers et al., 2001; Zou, 2014). Afterward, 1.6 g L$^{-1}$ of NaH$^{13}$CO$_3$ (isotope $^{13}$C 99% Aldrich) was added. H. scabra thalli fragments (2 g wet weight) were placed in hermetically sealed 250 mL BOD bottles ($n = 3$) containing seawater previously prepared with $^{13}$C isotope and maintained for 24 hours at 28°C of temperature under light saturation (278 µmol photons m$^{-2}$ s$^{-1}$, previously determined as H. scabra saturation irradiance, $I_k$ (ratio of $P_{\text{max}}/\alpha$, ...)
where $\alpha$ is photosynthetic efficiency. Three photoperiod treatments were selected: (1) 24 hours in light, (2) 12:12 hours light:darkness, and (3) 24 hours in darkness. A control bottle containing seawater without $^{13}$C isotope was used in each treatment. At the end of the incubations the algae were washed with seawater and rinsed with distilled water to remove the remains of the isotope that were not absorbed, and later frozen and lyophilized. Lyophilized samples (0.6 g) were depigmented twice in succession with methanol (100%) after that, low molecular weight carbohydrates were extracted in distilled water for 24 hours. Finally, the supernatant was frozen and lyophilized to be used in the NMR analysis.

$^{13}$C-Nuclear Magnetic Resonance Spectroscopy (NMR) analyses

To determine the incorporation of NaH$^{13}$CO$_3$ isotope in photosynthetic products, lyophilized samples (8 mg) were dissolved in 1 mL with 99.8% deuterium oxide (D$_2$O). The proton ($^{13}$C) spectra were recorded on a Varian/Agilent Premium Compact 600 NMR spectrometer (Palo Alto, CA, USA) at a frequency of 150.83 MHz using Sodium [3-trimethylsilyl 2,2',3,3'-2-H4] propionate (TSP-d4) with internal reference to 0.00 ppm. All NMR spectra were recorded at room temperature using the following parameters: scans = 50,000; $^{13}$C-pulse width of 3.3 s, an acquisition time of 0.5 s, and a relaxation delay of 0.60 s.

Statistical analyses

To test the interactive effects of temperature, pH, and nutrient levels on dependent variable ($P_{\text{gross}}$), a three-way ANOVA (3 x 3 x 4) was performed considering the three environmental factors as independent random variables. A Two-way ANOVA analyzed the effect of pH and nutrients on $Q_{10} P_{\text{gross}}$. One-way ANOVA was applied to test for differences between different inhibitor assays whereas Newman-Keuls post-hoc multiple comparisons were used to test between treatments. All statistical tests and analyses were performed using the statistical package Statistica™ 7. Before analyses, homogeneity of variance (Bartlett) and normality test (Kolmogorov-Smirnov) were tested, and transformations were applied if necessary, to meet such criteria.

RESULTS

Photosynthetic responses to the interactive effect of temperature, pH, and nutrient levels

The three-way ANOVA showed a significant interactive effect of temperature, pH, and nutrients on $H. scabra$ gross photosynthesis, $P_{\text{gross}}$ ($F_{12; 216} = 4.57, p \leq 0.001$) (Fig. 1; Table S2). The highest $P_{\text{gross}}$ values (1.83 mg O$_2$ g DW h$^{-1}$) were obtained at the highest nutrient concentration (10:1.0 μM) under elevated temperature (33°C) at a pH of 8.6 and in the control treatment (no added nutrients) at 33°C but, at the lowest pH (7.5) ($P_{\text{gross}} = 1.78$ mg O$_2$ g DW h$^{-1}$). In contrast, low and medium nutrient level treatments had lowest $P_{\text{gross}}$ at intermediate temperatures. In general, $H. scabra$ photosynthetic rates were higher at the highest temperature tested regardless of the nutrient or pH levels, except for the low nutrient treatment, which had higher $P_{\text{gross}}$ at the lowest temperature for all pH treatments.
It is noteworthy how $P_{\text{gross}}$ responds to temperature and pH changes in relation to nutrient concentrations in an opposite pattern between the highest nutrient concentration (10:1.0) and the control treatment (no added nutrients) at the highest temperature tested.

Individual analysis of factors showed that pH alone had no effect on $P_{\text{gross}}$ ($F = 2.84, p > 0.05$), whereas the individual effect of nutrients ($F = 4.77, p \leq 0.05$) and temperature ($F = 45.30, p \leq 0.001$) were significant (Table S2). All interactions involving two factors were significant: temperature-pH ($F = 2.70, p \leq 0.05$); nutrient-temperature ($F = 10.32, p \leq 0.001$); and pH-nutrients ($F = 9.23, p \leq 0.001$).

Effect of temperature on $P_{\text{gross}}$ ($Q_{10}$ $P_{\text{gross}}$) under nutrient concentrations and pH levels

The two-way ANOVA showed a significant effect of nutrient levels on $Q_{10}$ $P_{\text{gross}}$ ($F_{2, 54} = 6.721, p = 0.002$). This effect was more pronounced with the highest nutrient concentration ($Q_{10} = 2.42$) and decreased gradually as nutrient concentration decreased, from 1.75 to 0.75 in medium and low nutrient concentration, respectively. Conversely, pH and its interaction with nutrient levels did not show any significant effect on the $Q_{10}$ calculated values (Table 1).

HCO$_3^-$ Uptake

The $\delta^{13}$C value in $H. scabra$ was $-23.9\%$ suggesting uptake of both HCO$_3^-$ and CO$_2$, and the presence of a CCM. The carbonic anhydrase assays corroborate the latter since the addition of EZ caused a significant inhibition (22.2\%) of maximum photosynthesis rates $P_{\text{max}}$ ($F_{3, 24} = 18.674, p \leq 0.001$) whereas the combination of both inhibitors produced a similar effect to that found with EZ (Fig. 2). AZ inhibitor showed no effect on $P_{\text{max}}$ implying a lack of external CA and direct uptake of HCO$_3^-$ with a CCM depending on internal CA.

Incorporation of $^{13}$C into products of photosynthesis

$^{13}$C isotope labeling in $H. scabra$ observed by signal multiplicity (coupling) also showed bicarbonate uptake, since $^{13}$C isotope was incorporated into an amino acid akin to aspartate in the three photoperiod treatments analyzed. The incorporation in darkness indicates non-photosynthetic $\beta$-carboxylation. Aspartate also appears in the three control treatments (simple decoupled signal) highlighting its abundance in the species (Fig. 3).

DISCUSSION

The results of our work support the hypothesis that a synergistic increase in pH, temperature, and nutrients enhances $H. scabra$ photosynthesis. An increase in temperature could enhance $P_{\text{gross}}$ at high pH if there are sufficient nutrients. Environmental conditions of high seawater temperature (Robledo & Freile-Pelegrín, 2005; Rodríguez-Martínez et al., 2010), alkaline pH seawater, likely because of the karstic origins of the Yucatán Peninsula (Cejudo et al., 2020), and pulsatile nutrient enrichment due to the submarine groundwater discharge (Hernández-Terrones et al., 2015), are common in Quintana Roo coastal areas where $H. scabra$ and other $Halimeda$ species colonize shallow environments. Interestingly, in the control treatment (no added nutrients) at low pH and high temperature high $P_{\text{gross}}$ was also observed,
most probably because at low pH CO$_2$ availability is higher (Falkowski & Raven, 2007) and an increase in temperature facilitates photosynthesis (Bruno, Carr & O’Connor, 2015; Vásquez-Elizondo & Enríquez, 2016). Our results also emphasize that interactive effects are more reliable indicators than those observed under individual analyses since pH alone had no effect on $P_{gross}$ while all interactions were significant. Variation of only one factors can modify the photosynthesis response to other factors highlighting the importance of interactive studies.

Conversely, the interactive effect of decreasing pH (low and medium) with increases in temperature and nutrient enrichment kept $P_{gross}$ below its potential capacity. Thus, potential deleterious effects on $H$. scabra performance are expected to occur under future scenarios of ocean acidification, global warming, and their complex interactions with nutrient enrichment due to the continuous coastal development in the area.

In agreement with our results with $H$. scabra, significant reductions in gross photosynthetic rates have been reported for $H$. macroloba and $H$. cylindracea when exposed to elevated CO$_2$ combined with elevated temperature, showing an additive negative effect (Sinutok et al., 2012). In contrast, for $H$. incrassata, $H$. simulans, and $H$. opuntia no significant effects in net photosynthesis were reported for the interactions among species, pH, and temperature (Campbell et al., 2016). While, in $H$. opuntia no interactive effect of CO$_2$ and nutrient enrichment on net photosynthesis was found (Hofmann et al., 2015).

These contrasting results among congeners indicate that the photosynthetic responses to the interactive effects of several environmental variables are complex since, in addition to the factors being evaluated, the physiological mechanisms could be responding to other interrelated processes that were not assessed during assays. For example, Campbell et al. (2016) found in three Halimeda species that photosynthesis was positively correlated to calcification rates and, an increase in temperature increased activity of both processes. In this context, processes with high carbon requirements such as calcification could indirectly stimulate photosynthesis (Carvalho & Eyre, 2017), generating protons that are used to facilitate the absorption of nutrients and bicarbonate (McConnaughey & Whelan, 1997). The reduction of NO$_3^-$ to NH$_4^+$ is another process with high energy requirements (Ale, Mikkelsen & Meyer, 2011), and it is related to carbon fixation (Cabello-Pasini & Figueroa, 2005) so it is likely to be more plausible to affect photosynthesis rather than calcification since the latter appears to be more dependent on photosynthetic activity of many calcifying primary producers. Nutrient enrichment supported a rapid increase in the physiological performance of $H$. opuntia (Teichberg, Fricke & Bischof, 2013). Therefore, the photosynthetic increase found with the nutrient addition ($KNO_3$:K$_3$PO$_4$) in our experiments could be the result of its effect on processes related to nutrient uptake and these, to photosynthesis. Moreover, it is also known that nutrient uptake rates increase with temperature increases (Harrison & Hurd, 2001), consequently, our results not only are a response to the interactive effect of environmental factors but also, the result of the direct and indirect response of other metabolic processes on photosynthesis.

Temperature is a significant factor controlling metabolic rates, including photosynthesis; increasing temperature increases photosynthetic rates linearly up to an optimum rate, beyond this
thermal threshold rates, tend to decline (Bruno, Carr & O’Connor, 2015; Vásquez-Elizondo & Enríquez, 2016). It is generally accepted that $Q_{10}$ values greater than 2 characterize an active nutrient absorption process across cell membranes, while $Q_{10}$ ~ 1 values describe passive processes that are not greatly affected by temperature (Lobban & Harrison, 1994). According to this, our calculated $Q_{10}$ $P_{\text{gross}}$ for the medium and high nutrient treatments are in the range of active nutrient absorption, expected by organisms living in highly illuminated habitats and with high elevated metabolic activity (Vásquez-Elizondo & Enríquez, 2016).

Mechanisms of photosynthetic carbon uptake can influence the isotopic composition of organic matter. Values of $\delta^{13}C$ between -30 and -10‰ indicate both HCO$_3^-$ and CO$_2$ active uptake and species who fixation fall within this range are classified as species with active CCM (Maberly, Raven & Johnston, 1992; Raven et al., 2002; Diaz-Pulido et al., 2016; Bender-Champ, Diaz-Pulido & Dove, 2017). Species with $\delta^{13}C$ signatures between -32‰ and -22‰ are considered as C3 plants while $\delta^{13}C$ between –16‰ and –10‰ are typical for C4 plants (Rautenberger et al., 2015; Valiela et al., 2018). Considering these ranges and the results obtained in this work, *H. scabra* could be classified as a C3 plant with a CCM that uses both, HCO$_3^-$ and CO$_2$ as a resource of $C_i$ for photosynthesis. The $\delta^{13}C$ values of *H. scabra* found within this study are in the range of those reported in other Halimeda species, such as *H. opuntia*, ~-21‰ (Zweng, Koch & Bowes, 2018) and *H. tuna*, ~-21‰ (Duarte et al., 2018).

The extracellular CA inhibitor (AZ), did not show any adverse effect on photosynthesis, indicating a direct uptake of HCO$_3^-$ while a significant reduction in $P_{\text{max}}$ with EZ, confirmed HCO$_3^-$ uptake and the presence of a CCM (Badger et al., 1998) along with the role of internal CA (Badger & Price, 1994). The reduction of photosynthesis under the activity of EZ was only about 22.2% relative to control samples, likely because there was still enough CO$_2$ in the proximity of RuBisCO to maintain a reduced level of photosynthesis. The available CO$_2$ may come from the following alternatives: (1) as the result of a CCM, mainly related to efficient HCO$_3^-$ utilization (Raven, 1997); (2) respiratory CO$_2$ (Borowitzka & Larkum, 1976); (3) CO$_2$ supplied from interutricular spaces, although this source is not sufficient to sustain photosynthesis (De Beer & Larkum, 2001); and (4) by CO$_2$ diffusion from both, the external medium and the intercellular space (ICS) (Borowitzka & Larkum, 1976). All previous explanations could maintain RuBisCO CO$_2$ saturated and minimize photosynthetic losses after inhibiting intracellular CA. This suggests that inorganic carbon supply for photosynthesis in *H. scabra* does not depend entirely on the activity of the CA’s, and might be maintained by several mechanisms which may be advantageous during adverse conditions. Bicarbonate (HCO$_3^-$) uptake has also been found in *H. discoidea*, *H. macroloba*, and *H. tuna* (Borowitzka & Larkum, 1976), while the lack of extracellular CA has been found in *H. discoidea* (De Beer & Larkum 2001) and *H. cuneata* f. digitate (Hofmann et al., 2015a).

Some Halimeda species possess CCM and use bicarbonate as an alternate source of inorganic carbon (Borowitzka & Larkum, 1976; Price et al., 2011) keeping their photosynthesis carbon saturated under the current seawater $C_i$ concentrations (Beer, 1994; Beardall, Beer & Raven, 1998; Fernández, Hurd & Roleda, 2014). Therefore, this ability may also be responsible...
for $P_{\text{gross}}$ enhancement under elevated temperature observed in this study in *H. scabra* through a decrease in photorespiration (Ogren, 1984). According to Giordano, Beardall & Raven (2005), the number of resources that a cell invests in acquiring carbon through a CCM is likely to be coupled with the availability of nutrients. This could also explain the increase in $P_{\text{gross}}$ observed at the highest nutrient level since most CCM’s require the *de novo* synthesis of specific proteins, which represents a demand for cellular nitrogen (Giordano, Beardall & Raven, 2005). The low oxygen production observed at low pH (under high nutrient concentrations) in *H. scabra* could be a delay in the induction of the CCM relying on passive diffusion of CO$_2$ alone, thus leading to reduced efficiency of carbon assimilation (Price et al., 2011; Cornwall et al., 2012; Meyer et al., 2016).

In this study, *H. scabra* incorporated $^{13}$C isotope into aspartate in the three photoperiod treatments demonstrating photosynthetic and non-photosynthetic NaH$^{13}$CO$_3$ assimilation. Moreover, the incorporation of $^{13}$C isotope in aspartate in 24-hour darkness treatment indicates $\beta$-carboxylation which facilitates a metabolic alternative to inorganic carbon carboxylation resulting in an important contribution to CCMs (Raven & Osmond, 1992; Enríquez & Rodríguez-Román, 2006). $\beta$-carboxylation has multiple functions for algal metabolism such as providing essential compounds for growth that cannot be produced photosynthetically (Falkowski & Raven, 2007). Carbon fixation of these compounds can be done in both light and darkness (Axelsson, 1988), and it is generally less than 5% of maximum photosynthesis (Cabello-Pasini & Alberte, 1997). In marine algae, the end products of this carbon fixation independent of light are typically organic compounds and amino acids rather than triose sugars generated during photosynthesis (Cabello-Pasini & Alberte, 1997). Although the $\delta^{13}$C values in *H. scabra* suggest a C3 pathway, the abundance of aspartate in control and experimental treatments also suggest the existence of a C4 pathway. In C4 plants, the C4 acids malate and aspartate are the major initial photosynthetic products, these products are rapidly decarboxylated releasing CO$_2$ for its refixation by RuBisCO functioning as photosynthetic intermediates (Holaday & Bowes, 1980). In this sense, a C4 mechanism could explain the increase in $P_{\text{gross}}$ under all of our high-end treatments (temperature, pH, and nutrient concentration), as well as the insensitivity of the photosynthetic response of the alga to AZ and, its low inhibition in the presence of EZ. C4 plants have an active CCM, which is mainly related to the efficient use of HCO$_3^-$ through an initial carboxylation reaction by a Phosphoenolpyruvate enzyme (Badger & Price, 1994; Raven, 1997). C4 mechanisms have been reported in some Chlorophyta, including the semi-calcified *Udotea flabellum*, which shows an initial carboxylation by phosphoenolpyruvate carboxykinase enzyme (Reiskind, Seamon & Bowes, 1988), whereas in *Ulva prolifera* evidence of both C3 and C4 pathway has been found (Xu et al., 2012).

**CONCLUSIONS**

*H. scabra* uses both CO$_2$ and HCO$_3^-$ for photosynthesis and it seems to have different mechanisms for $C_i$ acquisition incorporating bicarbonate through photosynthetic and non-photosynthetic pathways. Our results suggest the presence of both C3 and C4 pathways, with the latter relying on $\beta$-carboxylation. These strategies give *H. scabra* physiological plasticity to...
acclimate to possible environmental changes in the short term. Our study strongly suggests that *H. scabra* acclimatizes better to environmental conditions of high pH and high temperature with enough nutrient enrichment. Although these conditions could exacerbate the presence of epiphytes and opportunistic algae, the availability of pulsatile nutrients likely plays a role in maintaining *Halimeda* populations by enhancing algal photosynthetic performance. Such conditions are typical in the Yucatan peninsula coast where *Halimeda* species grow in abundance. Opposite interactive conditions of decreasing pH in combination with increases in temperature and nutrient availability, could keep photosynthesis at a sub-optimal level which has strong ecological implications due to the potential for decline in *Halimeda* abundance and the resulting consequences to sediment production and carbon balance in coral reefs where these algae thrive.

**ACKNOWLEDGMENTS**

The authors thank Dr. D. Valdés, V. Avila Velázquez, and C. Chávez for their help in setting up the experiments and to Dr. E. Serviere-Zaragoza for the tissue isotopic analysis.

**REFERENCES**

Ale MT, Mikkelsen JD, Meyer AS. 2011. Differential growth response of *Ulva lactuca* to ammonium and nitrate assimilation. *Journal of Applied Phycology* 23:345–351 DOI: 10.1007/s10811-010-9546-2.

Alcolado PM, Claro-Madruga R, Menéndez-Macías G, Garcia-Parrado P, Martínez-Daranas B, Sosa M. 2003. The Cuban coral reefs. In: Cortés J, ed. Latin American Coral Reefs. Elsevier Science B.V. 56-76 DOI: 10.1016/B978-044451388-5/50004-7.

Álvarez-Cadena JN, Ordóñez-López U, Valdés-Lozano D, Almaral-Mendívil AR, Uicab-Sabido A. 2007. Estudio anual del zooplancton: composición, abundancia, biomasa e hidrología del norte de Quintana Roo, Mar Caribe de México. *Revista Mexicana de Biodiversidad* 78 (2):421-430.

Axelsson L. 1988. Changes in pH as a measure of photosynthesis by marine macroalgae. *Marine Biology* 97:287-294 DOI: 10.1007/BF00391314.

Badger MR, Andrews TJ, Whitney SM, Ludwig M, Yellowlees DC, Leggat W, Price GD. 1998. The diversity and coevolution of RuBisCO, plastids, pyrenoids, and chloroplast-based CO₂-concentrating mechanisms in algae. *Canadian Journal of Botany* 76:1052-1071 DOI: 10.1139/cjb-76-6-1052.

Badger MR, Price GD. 1994. The role of Carbonic Anhydrase in photosynthesis. *Annual Review of Plant Physiology and Plant Molecular Biology* 45:369-392 DOI: 10.1146/annurev.pp.45.060194.002101.

Bender-Champ D, Diaz-Pulido G, Dove S. 2017. Effects of elevated nutrients and CO₂ emission scenarios on three coral reef macroalgae. *Harmful Algae* 65: 40–51 DOI: 10.1016/j.hal.2017.04.004.
Beach K, Walters L, Borgeas H, Smith C, Coyer J, Vroom P. 2003. The impact of Dictyota spp. on Halimeda populations of Conch Reef, Florida Keys. *Journal of Experimental Marine Biology and Ecology* 297:141–159 DOI: 10.1016/j.jembe.2003.07.003.

Beardall J, Beer S, Raven JA. 1998. Biodiversity of marine plants in an era of climate change: some predictions based on physiological performance. *Botanica Marina* 41:113–123. DOI: https://doi.org/10.1515/botm.1998.41.1-6.113.

Beer S. 1994. Mechanisms of inorganic carbon acquisition in marine macroalgae (with special reference to the Chlorophyta). *Progress in phycological research* 10:179–207.

Bjork M, Haglund K, Ramazanov Z, Garcia-Reina G, Pedersen M. 1992. Inorganic carbon assimilation in the green seaweed *Ulva rigida* C. Ag. (Chlorophyta). *Planta* 187:152-156 DOI: 10.1007/BF00201637.

Borowitzka MA, Larkum AWD. 1976. Calcification in the green alga *Halimeda*. III. The sources of inorganic carbon for photosynthesis and calcification and a model of the mechanism of calcification. *Journal of Experimental Botany* 27: 879–893 DOI: 10.1093/jxb/27.5.879.

Bruno JF, Carr LA, O’Connor MI. 2015. Exploring the role of temperature in the ocean through metabolic scaling. *Ecology* 96(12):3126–3140 DOI: 10.1890/14-1954.1.

Cabello-Pasini A, Alberte RS. 1997. Seasonal patterns of photosynthesis and light-independent carbon fixation in marine macrophytes. *Journal of Phycology* 33:321-329 DOI: 10.1111/j.0022-3646.1997.00321.x.

Cabello-Pasini A, Figueroa FL. 2005. Effect of nitrate concentration on the relationship between photosynthetic oxygen evolution and electron transport rate in *Ulva rigida* (Chlorophyta). *Journal of Phycology* 41:1169–1177 DOI: 10.1111/j.1529-8817.2005.00144.x.

Campbell JE, Fisch J, Langdon C, Paul VJ. 2016. Increased temperature mitigates the effects of ocean acidification in calcified green algae (*Halimeda* spp.). *Coral Reefs* 35:357–368 DOI: 10.1007/s00338-015-1377-9.

Carvalho CM, Eyre DB. 2017. Light respiration by subtropical seaweeds. *Journal of Phycology* 53:486–492 DOI: 10.1111/jpy.12533.

Cejudo E, Acosta-González G, Ortega-Camacho D, Tun-Rosado GE. 2020. Changes in the hydrochemistry of a karstic lake in Yucatan, Mexico. *Environmental Earth Sciences* 79:98 DOI: 10.1007/s12665-020-8838-3.

Celis-Plá P, Hall-spencer JM, Horta P, Milazzo M, Korbee N, Cornwall CE, Figueroa FL. 2015. Macroalgal responses to ocean acidification depend on nutrient and light levels. *Frontiers in Marine Science* 2:26 DOI: 10.3389/fmars.2015.00026.

Cornwall C, Hepburn CD, Pritchard D, Currie KI, McGraw CM, Hunter KA, Hurd CL. 2012. Carbon-use strategies in macroalgae: differential responses to lowered pH and implications for ocean acidification. *Journal of Phycology* 48:137–144 DOI: 10.1111/j.1529-8817.2011.01085.x.
Cornwall CE, Revill AT, Hurd C. L. 2015. High prevalence of diffusive uptake of CO₂ by macroalgae in a temperate subtidal ecosystem. *Photosynthesis Research* 124:181–190. DOI: 10.1007/s11120-015-0114-0.

Crook ED, Potts D, Rebollodeo-Vieyra M, Hernández L, Paytan A. 2012. Calcifying coral abundance near low-pH springs: implications for future ocean acidification. *Coral Reefs* 31(1): 239-245. DOI 10.1007/s00338-011-0839-y.

De Beer D, Larkum AWD. 2001. Photosynthesis and calcification in the calcifying algae *Halimeda discoidea* studied with microsensors. *Plant Cell and Environment* 24:1209–1217. DOI: 10.1046/j.1365-3040.2001.00772.x.

Díaz-Pulido G, Cornwall C, Gartrell P, Hurd C, Tran DV. 2016. Strategies of dissolved inorganic carbon use in macroalgae across a gradient of terrestrial influence: implications for the Great Barrier Reef in the context of ocean acidification. *Coral Reefs* 35(4):1327-1341 DOI: 10.1007/s00338-016-1481-5.

Duarte CM, Delgado-Huertas A, Anton A, Carrillo-de-Albornoz P, López-Sandoval DC, Agustí S, Almahasheer H, Marbá N, Hendriks IE, Krause-Jensen D, Garcías-Bonet N. 2018. Stable Isotope (δ¹³C, δ¹⁵N, δ¹⁸O, δD) Composition and Nutrient Concentration of Red Sea Primary Producers. *Frontiers in Marine Science* 5:298 DOI: 10.3389/fmars.2018.00298.

Enríquez S, Rodríguez-Román A. 2006. Effect of water flow on the photosynthesis of three marine macrophytes from a fringing-reef lagoon. *Marine Ecology Progress Series* 323:119–132 DOI: 10.3354/meps323119.

Falkowski PG, Raven JA. 2007. Aquatic Photosynthesis: (Second Edition). Princeton; Oxford: Princeton University Press.

Fernández PA, Hurd CL, Roleda MY. 2014. Bicarbonate uptake via an anion exchange protein is the main mechanism of inorganic carbon acquisition by the giant kelp *Macrocystis pyrifera* (Laminariales, Phaeophyceae) under variable pH. *Journal of Phycology* 50:998–1008. https://doi.org/10.1111/jpy.12247.

Figueroa FL, Israel A, Neori A, Martínez B, Malta EJ, Ang P, Inken S, Marquardt R, Korbee N. 2009. Effects of nutrient supply on photosynthesis and pigmentation in *Ulva lactuca* (Chlorophyta): responses to short-term stress. *Aquatic Biology* 7:173–183 DOI: 10.3354/ab00187.

Giordano M, Beardall J, Raven JA. 2005. CO₂ Concentrating Mechanisms in Algae: Mechanisms, Environmental Modulation, and Evolution. *Annual Review of Plant Biology* 56:99–131. DOI: 10.1146/annurev.arplant.56.032604.144052.

Harley CDG, Anderson KM, Demes KW, Jorve JP., Kordas RL, Coyle TA, Graham MH. 2012. Effects of climate change on global seaweed communities. *Journal of Phycology* 48:1064–1078 DOI: 10.1111/j.1529-8817.2012.01224.x.

Harrison PJ, Hurd CL. 2001. Nutrient physiology of seaweeds: Application of concepts to aquaculture. *Cahiers De Biologie Marine* 42:71-82. DOI: oai:ecite.utas.edu.au:93212.
Hernández-Terrones ML, Kimberly AN, Null KA, Ortega-Camacho D, Paytan A. 2015. Water quality assessment in the Mexican Caribbean: Impacts on the coastal ecosystem. *Continental Shelf Research* 102: 62–72 DOI: 10.1016/j.csr.2015.04.015.

Hillis-Colinvaux L. 1980. Ecology and taxonomy of *Halimeda*: primary producer of coral reefs. *Advances in Marine Biology* 17:1-327 DOI: 10.1016/S0065-2881(08)60303-X.

Hofmann LC, Bischof K, Baggini C, Johnson A, Koop-Jakobsen K, Teichberg M. 2015. CO₂ and inorganic nutrient enrichment affect the performance of a calcifying green alga and its noncalcifying epiphyte. *Oecologia* 177(4):1157–1169 DOI: 10.1007/s00442-015-3242-5.

Hofmann L, Fink A, Bischof K, de Beer D. 2015a. Microsensor studies on *Padina* from a natural CO₂ seep: implications of morphology on acclimation to low pH. *Journal of Phycology* 51:1106–111 DOI: 10.1111/jpy.12347.

Hofmann LC, Heiden J, Bischof K, Teichberg M. 2014. Nutrient availability affects the response of the calcifying chlorophyte *Halimeda opuntia* (L.) J.V. Lamouroux to low pH. *Planta* 239(1):231-242 DOI:10.1007/s00425-013-1982-1.

Holaday AS, Bowes G. 1980. C4 Acid Metabolism and Dark CO₂ Fixation in a Submersed Aquatic Macrophyte (*Hydrilla verticillata*). *Plant Physiology* 65:331-335. DOI: 10.1104/pp.65.2.331.

Hurd CL, Hepburn CD, Currie KI, Raven JA, Hunter KA. 2009. Testing the effects of ocean acidification on algal metabolism: considerations for experimental designs. *Journal of Phycology* 45:1236–1251 DOI: 10.1111/j.1529-8817.2009.00768.x.

Invers O, Perez M, Romero J. 1999. Bicarbonate utilization in seagrass photosynthesis: role of carbonic anhydrase in *Posidonia oceanica* (L.) Delile and *Cymodocea nodosa* (Ucria) Ascherson. *Journal of Experimental Marine Biology and Ecology* 235:125–133 DOI: 10.1016/S0022-0981(98)00172-5.

Invers O, Zimmerman CR, Alberte SR, Perez A, Romero J. 2001. Inorganic carbon sources for seagrass photosynthesis: an experimental evaluation of bicarbonate use in species inhabiting temperate waters. *Journal of Experimental Marine Biology and Ecology* 265:203–217 DOI: 10.1016/S0022-0981(01)00332-X.

Kram SL, Price NN, Donham EM, Johnson MD, Kelly ELA, Hamilton SL, Smith JE. 2016. Variable responses of temperate calcified and fleshy macroalgae to elevated pCO₂ and warming. *ICES Journal of Marine Science* 73(3):693–703 DOI:10.1093/icesjms/fsv168.

Lignell Å, Pedersén M. 1989. Effects of pH and inorganic carbon concentration on growth of *Gracilaria secundata*. *British Phycological Journal* 24:83-89, DOI: 10.1080/00071618900650071.

Lobban CS, Harrison PJ. 1994. Seaweed Ecology and Physiology. Press Syndicate of the University of Cambridge, Cambridge. 366 pp.

Maberly S, Raven J, Johnston A. 1992. Discrimination between¹²C and¹³C by marine plants. *Oecologia* 91(4):481–492 DOI: 10.1007/BF00650320.
McConnaughey TA, Whelan JF. 1997. Calcification generates protons for nutrient and bicarbonate uptake. *Earth-Science Reviews* 42:95-117 DOI: 10.1016/S0012-8252(96)00036-0.

Meyer FW, Schubert N, Diele K, Teichberg M, Wild C, Enríquez S. 2016. Effect of inorganic and organic carbon enrichments (DIC and DOC) on the photosynthesis and calcification rates of two calcifying green algae from a Caribbean reef lagoon. *Plos One* 11(8):e0160268 DOI: 10.1371/journal.pone.0160268.

Ogren WL. 1984. Photorespiration: pathways, regulation, and modification. *Annual Review of Plant Physiology* 35(1):415–442 DOI: 10.1146/annurev.pp.35.060184.002215.

Peach KE, Koch MS, Blackwelder PL. 2016. Effects of elevated pCO$_2$ and irradiance on growth, photosynthesis and calcification in *Halimeda discoidea*. *Marine Ecology Progress Series* 544:143-158 DOI: 10.3354/meps11591.

Price NN, Hamilton SL, Tootell SJ, Smith EJ. 2011. Species-specific consequences of ocean acidification for the calcareous tropical green algae *Halimeda*. *Marine Ecology Progress Series* 440:67–78 DOI:10.3354/meps09309.

Raven JA. 1997. Inorganic carbon acquisition by marine autotrophs. *Advances in Botanical Research* 27:85–209. https://doi.org/10.1016/S0065-2296(08)60281-5.

Raven JA, Osmond CB. 1992. Inorganic C assimilation processes and their ecological significance in inter- and sub-tidal macroalgae of North Carolina. *Functional Ecology* 6:41–47 DOI:10.2307/2389769.

Raven JA, Johnston AM, Kubler JE, Korb R, McNroy SG, Handley LL, Scrimgeour CM, Walker DI, Beardall J, Vanderklift M, Fredriksen S, Dunton KH. 2002. Mechanistic interpretation of carbon isotope discrimination by marine macroalgae and seagrasses. *Functional Plant Biology* 29(3):355–378 DOI: 10.1071/pp01201.

Raven JA, Hurd CL. 2012. Ecophysiology of photosynthesis in macroalgae. *Photosynthesis Research* 113:105–125 DOI: 10.1007/s11120-012-9768-z.

Robledo D, Freile-Pelegrín Y. 2005. Seasonal variation in photosynthesis and biochemical composition of *Caulerpa* spp. (Bryopsidales, Chlorophyta) from the Gulf of Mexico. *Phycologia* 44(3):312–319. DOI:10.2216/0031-8884(2005)44[312:SVIPAB]2.0.CO;2.

Rodríguez-Martínez RE, Ruiz-Rentería F, van Tussenbroek B, Barba-Santos G, Escalante-Mancera E, Jordán-Garza G, Jordán-Dahlgren E. 2010. Environmental state and tendencies of the Puerto Morelos CARICOMP site, Mexico. *The Revista de Biologia Tropical* 58(3):23-43 DOI: S0034-77442010000700004.

Sinutok S, Hill R, Doblin MA, Kühl M, Ralph PJ. 2012. Microenvironmental changes support evidence of photosynthesis and calcification inhibition in *Halimeda* under ocean acidification and warming. *Coral Reefs* 31:1201–1213 DOI: 10.1007/s00338-012-0952-6.
Stengel DB, Conde-Álvarez R, Connan S, Nitschke U, Arenas F, Abreu H, Bonomi Barufi J, Chow F, Robledo D, Malta EJ, Mata M, Konotchkick T, Nassar C, Pérez-Ruzafa Á, López D, Marquardt R, Vaz-Pinto F, Celis-Plá PSM, Hermoso M, Ruiz E, Ordoñez G, Flores P, Zanolla M, Bañares-España E, Altamirano M, Korbee N, Bischof K, Figueroa FL. 2014. Short-term effects of CO₂, nutrients and temperature on three marine macroalgae under solar radiation. *Aquatic Biology* 22:159–176 DOI: 10.3354/ab00576.

Teichberg M, Fricke A, Bischof K. 2013. Increased physiological performance of the calcifying green macroalga *Halimeda opuntia* in response to experimental nutrient enrichment on a Caribbean coral reef. *Aquatic Botany* 104:25-33 DOI: 10.1016/j.aquabot.2012.09.010.

Thomas MLH. 1988. Photosynthesis and respiration of aquatic macro-flora using the light and dark bottle oxygen method and dissolved oxygen analyzer. In: Lobban CS, Chapman DJ, Kremer BP, eds. Experimental Phycology: A Laboratory Manual. Cambridge University Press, Cambridge: 64-77.

Valiela I, Liu D, Lloret K, Chenoweth K, Hanacek D. 2018. Stable isotopic evidence of nitrogen sources and C4 metabolism driving the world’s largest macroalgal green tides in the Yellow Sea. *Scientific Reports* 8:17437 DOI: 10.1038/s41598-018-35309-3.

Vásquez-Elizondo RM, Enríquez S. 2016. Coralline algal physiology is more adversely affected by elevated temperature than reduced pH. *Scientific Reports* 6:1-14 DOI: 10.1038/srep19030.

Vogel N, Meyer FW, Wild C, Uthicke S. 2015. Decreased light availability can amplify the negative impacts of ocean acidification on calcifying coral reef organisms. *Marine Ecology-Progress Series* 521:49–61 DOI: 10.3354/meps11088.

Wernberg T, Bettignies T, Joy BA, Finnegan PM. 2016. Physiological responses of habitat-forming seaweeds to increasing temperatures. *Limnology and Oceanography* 61(6):2180-2190 DOI: 10.1002/lno.10362.

Xu J, Fan X, Zhang X, Xu D, Mou S, Cao S, Zheng Z, Miao J, Ye N. 2012. Evidence of Coexistence of C3 and C4 Photosynthetic Pathways in a Green-Tide-Forming Alga, *Ulva prolifera*. *PloS One* 7(5):e37438 DOI: 10.1371/journal.pone.0037438.

Zar JH. 1996. Biostatistical analysis. Prentice-Hall, Eryelwood Cliffs, N.J. 663 pp. Printed.

Zou D. 2014. The effects of severe carbon limitation on the green seaweed, *Ulva conglobata* (Chlorophyta). *Journal of Applied Phycology* 26:2417–2424 DOI: 10.1007/s10811-014-0268-8.

Zubia M, Freile-Pelegrín Y, Robledo D. 2014. Photosynthesis, pigment composition and antioxidant defenses in the red alga *Gracilariopsis tenuifrons* (Gracilariales, Rhodophyta) under environmental stress. *Journal of Applied Phycology* 26:2001-2010 DOI: 10.1007/s10811-014-0253-3.

Zweng RC, Koch MS, Bowes G. 2018. The role of irradiance and C-use strategies in tropical macroalgae photosynthetic response to ocean acidification. *Scientific Reports* 8:9479 DOI: 10.1038/s41598-018-27333-0.
Figure legends

Figure 1. Interactive effect of temperature, pH, and nutrients on the gross photosynthesis rate \( P_{\text{gross}} \) of \( H. \text{scabra} \) \( (n = 7) \). Symbols represent the mean and error bars 0.95 confidence intervals.

Figure 2. Comparison of the effect of two carbonic anhydrase inhibitors on \( P_{\text{max}} \) percentage. Error bars represent confidence intervals at 0.95. \( (n = 7) \).

Figure 3. NMR Spectra of NaH\(^{13}\)CO\(_3\) incorporation into photosynthetic products of \( H. \text{scabra} \) in different light and darkness treatments \( (n = 3) \). A) 24 h at saturation irradiances; B) 12 hours under light saturation and 12 h in darkness and, C) 24 h in darkness. * Indicate consistent signals for aspartate in control treatment; ** Indicate \(^{13}\)C enrichment (multiple coupling). Letter a indicates control, Letter b indicates treatment.
Figure 1

Effect of temperature, pH, and nutrients on the gross photosynthesis rate ($P_{\text{gross}}$) of *H. scabra*

Interactive effect of temperature, pH, and nutrients on the gross photosynthesis rate ($P_{\text{gross}}$) of *H. scabra* ($n = 7$). Symbols represent the mean and error bars 0.95 confidence intervals.
Figure 2

Carbonic anhydrase inhibitors on *H. scabra* $P_{\text{max}}$

Comparison of the effect of two carbonic anhydrase inhibitors on $P_{\text{max}}$ percentage. Error bars represent confidence intervals at 0.95. ($n = 7$).
Figure 3

NMR Spectra of NaH$_{13}$CO$_3$ incorporation into photosynthetic products of *H. scabra* in different light and darkness treatments (*n* = 3).

A) 24 h at saturation irradiances; B) 12 hours under light saturation and 12 h in darkness and, C) 24 h in darkness. * Indicate consistent signals for aspartate in control treatment; ** Indicate $^{13}$C enrichment (multiple coupling). Letter *a* indicates control, Letter *b* indicates treatment.
**Table 1 (on next page)**

Effect of temperature on $P_{\text{gross}} (Q_{10})$ in *H. scabra*.

Effect of temperature on $P_{\text{gross}} (Q_{10})$ in three nutrient concentrations and three pH levels (Two Way-ANOVA).
Table 1:
Effect of temperature on $P_{\text{grass}} (Q_{10})$ in three nutrient concentrations and three pH levels (Two Way-ANOVA).

|                          | Mean $Q_{10}$ | SS  | DF  | MS  | F    | p    |
|--------------------------|---------------|-----|-----|-----|------|------|
| Nutrient ratio           |               |     |     |     |      |      |
| 1.0:0.5                  | 10.839        |     | 2   | 5.419 | 6.721 | 0.002*** |
| 5.0:0.5                  | 0.75$^b$      |     |     |     |      |      |
| 10.0:1.0                 | 1.75$^a$      |     |     |     |      |      |
| pH                       | 3.137         |     | 2   | 1.568 | 1.945 | 0.152 ns |
| nutrients*pH             | 6.111         |     | 4   | 1.528 | 1.895 | 0.125 ns |
| Error                    | 43.543        |     | 54  | 0.806 |      |      |

ns: not significant; *** significant; letters indicate significant differences among mean based on Neuman-Keuls post hoc test.