Unveiling the Toxicity of Fine and Nano-Sized Airborne Particles Generated from Industrial Thermal Spraying Processes in Human Alveolar Epithelial Cells

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Abstract: High-energy industrial processes have been associated with particle release into workplace air that can adversely affect workers’ health. The present study assessed the toxicity of incidental fine (PGFP) and nanoparticles (PGNP) emitted from atmospheric plasma (APS) and high-velocity oxy-fuel (HVOF) thermal spraying. Lactate dehydrogenase (LDH) release, 2-(4-nitrophenyl)-2H-tetrazolium-1,3-benzene disulfonate (WST-1) metabolism, intracellular reactive oxygen species (ROS) levels, cell cycle changes, histone H2AX phosphorylation (γ-H2AX) and DNA damage were evaluated in human alveolar epithelial cells at 24 h after exposure. Overall, HVOF particles were the most cytotoxic to human alveolar cells, with cell viability half-maximal inhibitory concentration (IC50) values of 20.18 µg/cm² and 1.79 µg/cm² for PGFP and PGNP, respectively. Only the highest tested concentration of APS-PGFP caused a slight decrease in cell viability. Particle uptake, cell cycle arrest at S + G2/M and γ-H2AX augmentation were observed after exposure to all tested particles. However, higher levels of γ-H2AX were found in cells exposed to APS-derived particles (~16%), while cells exposed to HVOF particles exhibited increased levels of oxidative damage (~17% tail intensity) and ROS (~184%). Accordingly, APS and HVOF particles seem to exert their genotoxic effects by different mechanisms, highlighting that the health risks of these process-generated particles at industrial settings should not be underestimated.

Keywords: A549 cells; cell cycle; cytotoxicity; DNA damage; in vitro toxicity; incidental nanoparticles; H2AX phosphorylation; occupational exposure; process-generated nanoparticles

1. Introduction

The ceramic industry has been benefitting from nanotechnology innovation processes and advanced materials [1]. Workers from these industries are at risk of exposure to airborne fine (<2.5 µm mass median aerodynamic diameter (MMAD)) and nano-sized (<0.200 µm MMAD) particles that may be released either during the handling or manufacturing of
ceramics using engineered nanoparticles (ENP, 1–100 nm) as raw materials or to incidentally emitted particles during mechanical and combustion/heating processes. Indeed, high-energy processes such as laser ablation, laser sintering, physical vapour deposition, inkjet printing, thermal spraying processes (e.g., atmospheric plasma spraying (APS) and high-velocity oxy-fuel spraying (HVOF)) and glazing represent a high potential of fine and ultrafine particle formation and release to the workplace air [2–10].

Inhalation is the predominant route of exposure to micro- and nano-sized particles in occupational settings. Respiratory tract deposition and clearance is governed by aerosol physics and by the anatomy and physiology of the respiratory tract [11]. The deeper the particles reach, the harder their removal from the respiratory system, favouring particle–cell interactions that might result in adverse health effects [12,13]. Many epidemiological studies already described an association between exposure to particulate matter (PM) and the occurrence of adverse health effects [14,15]. Exposure to airborne particles has been associated with cardiovascular, pulmonary and neurological diseases, which leads to an increased risk of mortality [16–18]. In the case of the ceramic industry, worker exposure to ceramic dusts has been strongly linked to respiratory symptoms, such as wheezing, breathlessness and dry cough, as well as with reduced lung function, chronic bronchitis and chronic obstructive pulmonary disease (COPD) [19–21].

Most of the knowledge on nano(particle) toxicity comes from in vitro mechanistic studies. While there are many available studies on the in vitro toxicity of engineered nanoparticles (ENP), little is known about the in vitro hazard of process-generated (nano)particles. In this regard, our group recently conducted a comparative assessment of the in vitro toxicity of ENP (tin oxide (Sb₂O₃) and zirconium oxide (ZrO₂) ENP) that are used as input materials in the ceramic industry vs. particles collected during the HVOF spraying of ceramic coatings onto metal surfaces to produce thermal-resistant coatings [22]. Overall, our data showed that human tri-dimensional (3D) bronchial cultures under air–liquid interface (ALI) conditions were rather resistant to the ENP that induced mild cytotoxicity at early timepoints (24 h), though cells rapidly recovered since no significant changes in cell viability compared to the control were observed at late timepoints (72 h). At the same time, while the fine fraction of the HVOF-derived particles significantly decreased cell viability, the ultrafine fraction significantly increased DNA oxidative damage, showing that HVOF particles exhibited a higher toxicity potential compared to ENP [22]. A recent study by Cediel-Ulloa et al. also evaluated the in vitro toxicity of airborne particles emitted during gas–metal arc welding (GMAW) in a laboratory setting on primary human small airway epithelial cells (hSAEC) [23]. These authors observed that stainless steel welding particles were more cytotoxic compared to mild steel particles and induced oxidative stress in primary human small airway epithelial cells. In addition, Pavlovska et al. investigated the biological effects of airborne particulates collected in woodworking and metalworking industries both on EpiAirway 3D human small airway epithelial cells exposed for 4 h (half a working day), 8 h (full working day) and 72 h (three working days) and on A549 lung epithelial cells that were continuously exposed for 96 h. The obtained data showed that exposure to these polluting particles exerted minor acute effects on the morphology and viability of both A549 cells and EpiAirway tissues. However, a marked reduction in EpiAirway tissue viability after 8 h of exposure to woodworking particles and a slight reduction in tissue viability after 72 h of exposure to metalworking particles were observed [24].

Therefore, the present study aims to further explore the in vitro toxicity of process-generated fine (PGFP; <2.5 µm MMAD) and nano-sized particles (PGNP; <0.200 µm MMAD) that are incidentally emitted during two industrial thermal spraying processes (APS and HVOF) of ceramic coatings onto metal surfaces. We hypothesise that PGNP are more toxic to human alveolar epithelial cells than PGFP. Biological endpoints, including particle internalisation, plasma membrane integrity, cell metabolic activity and viability, reactive oxygen-species (ROS) levels, cell cycle analysis, histone H2AX phosphorylation and DNA damage, were evaluated in human alveolar epithelial A549 cells at 24 h after exposure.
2. Results
2.1. Process-Generated Fine and Nano-Sized Particle Characterisation

Table 1 presents the physicochemical features of the aqueous suspensions of the tested process-generated particles, namely, concentration (both in terms of mass and number of particles per mL), hydrodynamic size and oxidative potential. As shown, APS-derived aqueous suspensions were more diluted in terms of mass/mL than the HVOF ones, which somehow limited the maximum tested concentrations of the former type of particles, in particular of the PGNP fraction. For APS-derived particles, the mean hydrodynamic size value was 244 nm and 410 nm for PGFP and PGNP, respectively. At the same time, HVOF-PGFP (247 nm) and PGNP (236 nm) exhibited a similar hydrodynamic size mean value. Regarding the oxidative potential, only APS-derived PGFP demonstrated a low oxidative potential, whereas HVOF-derived particles, particularly PGNP, exhibited a high ability to produce •OH in a cell-free environment.

Table 1. Physicochemical characteristics of the tested PGFP and PGNP aqueous suspensions.

| Stock Suspension | Concentration (mg/mL) | Concentration (Number of Particles/mL) | Hydrodynamic Size (nm) | Oxidative Potential (A.U.) * |
|------------------|-----------------------|----------------------------------------|------------------------|----------------------------|
| APS              |                       |                                        |                        |                            |
| PGFP             | 0.068                 | 8.49 × 10^8                           | 244 ± 120              | 3291                       |
| PGNP             | 0.034                 | 4.21 × 10^8                           | 410 ± 162              | 5319                       |
| HVOF             |                       |                                        |                        |                            |
| PGFP             | 1.069                 | 9.72 × 10^8                           | 247 ± 116              | 9893                       |
| PGNP             | 0.140                 | 15.86 × 10^8                          | 236 ± 86               | 12833                      |

Data are presented as means ± SD. Hydrodynamic size and concentration were measured by nanoparticle tracking analysis. Oxidative potential was measured by electronic spin resonance. A.U.: arbitrary units. * Negative control (ultrapure water) = 3191 A.U.; Positive control (DOFA) = 48,041 A.U.

2.2. Plasma Membrane Integrity and Cell Viability

Under our experimental conditions, no differences in cell membrane integrity and viability were found after 24 h of exposure to both fractions of APS-derived particles compared with the negative control (NC) (Figure 1A). On the other hand, the cytotoxic effects were more pronounced when cells were exposed to particles derived from the HVOF spraying process. While a significant increase in the LDH release compared to control cells was observed after exposure to PGFP at either 10.00 µg/cm² (30.70%) or 20 µg/cm² (26.15%), a clear concentration-dependent decrease in cell viability was detected (Figure 1B). The analysis of cell viability concentration–response curves of cells exposed to HVOF-derived PGFP and PGNP revealed half-maximal inhibitory concentrations (IC₅₀) of 20.18 µg/cm² (CI 95%: 11.66–34.95) and 1.79 µg/cm² (CI 95%: 1.48–2.16), respectively.

2.3. Intracellular Reactive Oxygen Species Levels

No major effects on ROS levels were found on human alveolar epithelial cells after 24 h of exposure to APS-derived PGFP and PGNP (Figure 2A). On the other hand, in cells exposed to the HVOF particles, a significant increase in the intracellular ROS levels was found for the highest tested concentrations of PGFP (176.88 ± 32.35%) and PGNP (183.67 ± 59.06%), when compared to the NC (Figure 2B).

2.4. Cellular Uptake of the (Nano)Particles

The cellular uptake of the process-generated particles under study by human alveolar epithelial cells was estimated based on changes in the percentage of the side scatter signal (%SSC), a measure of cellular complexity, analysed by flow cytometry. As depicted in Figure 3, a concentration-dependent increase in cell complexity was observed in A549 cells that were incubated with all tested particles, regardless of the process and the particle fraction. For APS- and HVOF-derived particles, a significant increase in particle uptake was detected in cells exposed to the highest tested concentrations of either PGFP or PGNP. However, the fine HVOF-derived particles were internalised by A549 cells to a higher degree.
than the nano-sized fraction. At the same time, HVOF-PGFP (5 µg/cm²; 4.16 ± 1.61%) were internalised to a greater extent than APS-PGFP (5 µg/cm²; 2.69 ± 0.75%) in A549 cells.

Figure 1. Cytotoxicity of PGFP and PGNP released during APS (A) and HVOF (B) in human alveolar epithelial cells after 24 h of exposure. Data are expressed as means ± standard deviations (n = 3–4). LDH release values were normalised considering the positive control (total LDH release; cells lysed with 2% Triton X-100), while WST-1 reduction values were normalised considering the negative control (NC). Data were analysed by the one-way ANOVA test followed by Dunnett’s post hoc test for multiple comparisons. * p ≤ 0.05, ** p ≤ 0.01 and *** p ≤ 0.001 vs. NC. PC: Positive control (LDH: 2% Triton X-100; WST-1: 70% EtOH).

Figure 2. ROS intracellular levels in human alveolar epithelial cells after 24 h of exposure to PGFP and PGNP released during APS (A) and HVOF (B). Data are expressed as means ± standard deviations (n = 3–4). Values were normalised considering the NC. Data were analysed by the one-way ANOVA test followed by Dunnett’s post hoc test for multiple comparisons. * p ≤ 0.05, ** p ≤ 0.01 and *** p ≤ 0.001 vs. NC. PC: 25 µM AgNO₃.
with this increase being more marked in cells incubated with the APS-derived particles (Figure 5A) compared to the ones exposed to the HVOF particles (Figure 5B). In fact, for APS, increasing levels of γ-H2AX were found for each phase of the cell cycle after 24 h of exposure to the highest tested concentration of PGFP and PGNP APS-derived particles. Regarding HVOF-derived particles, increasing levels of γ-H2AX in each phase of the cell cycle were found but to a much lower degree than APS particles. Camptothecin (3.5 µg/mL) served as a positive control (PC) and, as expected, caused an evident increment in the total γ-H2AX levels in cells at the different cell cycle phases compared to the NC.

Representative graphs of the cellular uptake of particles, cell cycle and γ-H2AX analyses by flow cytometry are depicted in Figure A1.
2.5. Cell Cycle Analysis

Figure 4 shows the effect of the tested process-generated particles on the human alveolar epithelial cell cycle dynamics, as assessed by flow cytometry. In the NC, most of the cells were at the G₀/G₁ phase (78.04 ± 2.51% cells), which was expected, considering that the cells were incubated for 24 h in FBS-free culture medium. Moreover, a concentration-dependent increase in cells in the S and G₂/M phases was observed at 24 h after exposure to all particles under study, but G₀/G₁ cells still represented the largest subpopulation (42.31 ± 4.89% of the cells). Data were analysed by one-way ANOVA followed by Dunnett’s post-hoc test. * \( p \leq 0.05 \), ** \( p \leq 0.01 \) and *** \( p \leq 0.001 \) vs. NC.

2.7. Primary and Oxidative DNA Damage

Figure 5. \( \gamma \)-H2AX in human alveolar epithelial cells after 24 h of exposure to PGFP and PGNP released during APS (A) and HVOF (B). The percentage of cells in the sub-G₁ phase (apoptotic cells) was also analysed (C). Data were expressed as means ± standard deviations (n = 3–4). Data were analysed by one-way ANOVA followed by Dunnett’s post-hoc test. Global \( \gamma \)-H2AX analysis: $\$\leq 0.01$ and $\$\leq 0.001$ vs. NC. \( \gamma \)-H2AX in each phase of cell cycle: * \( p \leq 0.05 \), ** \( p \leq 0.01 \) and *** \( p \leq 0.001 \) vs. NC.
2.7. Primary and Oxidative DNA Damage

Figure 6 depicts the primary and oxidative DNA damage of A549 cells incubated with the APS- (Figure 6A) and HVOF-derived particles (Figure 6B). Although a slight increase in DNA strand breaks was observed for PGFP of both APS and HVOF spraying processes as well as for HVOF-derived PGNP, those were not significant when compared to the NC. However, APS-derived PGFP significantly increased DNA FPG-sensitive sites of human alveolar epithelial cells at concentrations of 2.5 $\mu$g/cm$^2$ (7.81 ± 4.40% tail intensity) and 5 $\mu$g/cm$^2$ (8.37 ± 2.23% tail intensity) compared to the NC (2.90 ± 1.80% tail intensity) (Figure 6A), while the nano-sized fraction did not increase oxidative DNA damage in A549 cells (Figure 6A). HVOF particles seem to cause higher levels of DNA oxidative damage compared to the APS particles. As shown in Figure 6B, both the PGFP and PGNP HVOF-derived fractions significantly increased DNA oxidation in a concentration-dependent manner. Representative comet images of the DNA damage of human alveolar epithelial cells exposed to the highest tested concentrations of PGFP and PGNP emitted during the HVOF thermal spraying process are depicted in Figure A2.

![Graphs showing DNA damage](image)

**Figure 6.** Primary (DNA strand breaks) and oxidative (FPG-sensitive) DNA damage in human alveolar epithelial cells after 24 h of exposure to PGFP and PGNP particles released during APS (A) and HVOF (B). Data are expressed as means ± standard deviations ($n = 3–4$). Data were analysed by one-way ANOVA followed by Dunnett’s post-hoc test. * $p \leq 0.05$ and *** $p \leq 0.001$ vs. NC. PC: 500 $\mu$M MMS and 2.5 mM KBrO$_3$ for primary and oxidative DNA damage, respectively.

3. Discussion

Regardless the size fraction, HVOF-derived particles were more cytotoxic for A549 cultures than the APS particles. Indeed, particles emitted during HVOF spraying induced a marked decrease in cell viability, with PGNP being more potent than PGFP, as evidenced by its $\sim 10 \times$ lower IC$_{50}$ value, while only the fine fraction of APS-derived particles slightly decreased cell viability at the highest tested concentration (5 $\mu$g/cm$^2$). In addition, only exposure to HVOF- but not to APS-emitted particles significantly increased the ROS intra-
cellular levels of A549 cells. These results are in good agreement with the highest oxidative potential of HVOF particles compared to the APS particles. ROS are important molecules that are involved in redox-associated signalling pathways. While important for regulating and maintaining normal physiological functions, excessive levels of intracellular ROS can activate cell death and other signalling pathways, including nuclear factor-κB (NF-κB), mitogen-activated protein kinase (MAPK) and phosphoinositide 3-kinase (PI3-K), that are involved in the regulation of the expression of inflammatory response genes, cell cycle arrest, DNA strand breaks and the formation of 8-oxo-7,8-dihydro-2′-deoxyguanosine (8-oxo-dG) DNA adducts [25,26].

Differences in their chemical composition might have an obvious toxicological impact and may explain the observed cellular effects in A549 cells. The chemical analysis of the airborne fine and nano-sized particles under study revealed a major enrichment in potentially health hazardous metals (chromium (Cr), nickel (Ni) and tungsten (W)) sourcing directly from the feedstock in both scenarios as well as in major elements (aluminium (Al), calcium (Ca) and iron (Fe)) with different possible source origins, including re-suspension of indoor dusts. Size-resolved particle chemical composition analysis by inductively coupled plasma mass spectrometry (ICP-MS) and inductively coupled plasma-optical emission spectrometry (ICP-OES) showed that APS-derived particles were mainly constituted by Al (68% and 42% for PGFP and PGNP, respectively), Cr (11% and 16% for PGFP and PGNP, respectively) and Ni (1% and 21% for PGFP and PGNP, respectively), while the HVOF-generated ones were mainly constituted by Cr (61% and 67% for PGFP and PGNP, respectively) and Ni (27% and 28% for PGFP and PGNP, respectively), as previously reported [2]. These differences in composition might have contributed to the particle aggregation/agglomeration of the aqueous suspensions, in particular of APS-PGNP that presented a high hydrodynamic size value.

Epidemiological and occupational studies have shown the hazard of inhalation exposure to Al, Cr and Ni to human health. Workplace exposure to airborne particles containing these elements has been associated with several respiratory disorders, such as pulmonary fibrosis, asthma, chronic obstructive lung disease, lung cancer and cardiovascular disease [27–34]. Several in vivo and in vitro studies have also addressed the toxicity of the nanoforms of these elements. Kim et al. investigated the toxicity of aluminium oxide (Al$_2$O$_3$) nanoparticles (NP) following 28 days of repeated exposure by inhalation in male rats and reported a no-observed-adverse-effect level (NOAEL) of 1 mg/m$^3$ [35]. At the same time, the existing evidence on the Al$_2$O$_3$ NP effects on human pulmonary cell lines seems to point out the minimal toxic effects caused by these NP that were considered less toxic when compared to cerium oxide (CeO$_2$), titanium dioxide (TiO$_2$), silicon dioxide (SiO$_2$) and zinc oxide (ZnO) NP [36–39]. These findings are in line with our results showing that Al-rich APS-derived particles did not induce a marked cytotoxicity or increment in the ROS levels of A549 cells.

Both tested thermal-spraying-derived particles, especially HVOF-emitted ones, are enriched in Cr and Ni. So far, the available in vitro and in vivo studies on Cr and Ni effects have shown pulmonary toxicity in response to exposure to these elements [40–45]. In this regard, human lung cells incubated with Cr(VI) have been reported to exhibit significant levels of oxidative DNA damage [46] as well as increasing levels of H2AX-Ser139 phosphorylation [47]. Evidence of cell cycle arrest at G$_2$/M has also been found in alveolar A549 cells exposed for 24 h to Cr(VI) [48]. On the other hand, Ding et al. showed that exposure to Ni triggered G$_2$/M cell cycle arrest and proliferation blockage in human bronchial epithelial cells (BEAS-2B) [49]. In the nanoscale form, DNA damage accompanied by increased phosphorylation of DNA damage response-associated proteins ATM serine/threonine protein kinase (Ser1981), p53 tumour protein (Ser15) and H2AX (ser139) has been reported by Mo et al. in BEAS-2B cells exposed to Ni NP for 24 h [50].

In response to DNA damage, the cell goes through various checkpoint mechanisms that are essential to survival, but when these fail, rapid cell death is a potential result [51]. DNA damage occurring throughout interphase will elicit a cell cycle arrest that allows
time for repair mechanisms to occur prior to progression to the subsequent phases of the cell cycle [52]. For instance, depending on the cell cycle phase at which a double strand break occurs, the repair mechanism used by the cell differs [53]. One mechanism of double strand break repair is through the phosphorylation-dependent recruitment of DNA damage repair factors to sites of DNA damage, such as the phosphorylation of Ser139 on histone H2AX (γ-H2AX) [54]. In the current study, flow cytometry data showed that exposure to any type of thermal-spraying-derived particles induced cell cycle arrest at the S and G₂/M phases in human alveolar epithelial A549 cells, most likely triggered in response to DNA damage. Indeed, following DNA damage, the S-phase checkpoint delays DNA synthesis, while the G₂ cell cycle checkpoint prevents cells from entering mitosis, inhibiting cell proliferation [52]. In addition, for every particle, a concentration-dependent increase in the number of apoptotic sub-G₁ phase cells after 24 h of exposure was observed. Under our experimental conditions, an overall increase in γ-H2AX levels was found in A549 cells, especially in cells at the S and G₂/M phases. For both types of process-generated particles, PGNP were more effective in increasing γ-H2AX levels compared to PGFP. Interestingly, a more prominent effect on γ-H2AX levels was observed in cells incubated with fine or nano-sized APS-derived particles. On the other hand, although both fractions of HVOF-emitted particles increased the oxidative DNA damage, in cells exposed to PGNP, that effect was visible at lower concentrations. At the same time, oxidative DNA lesions were only observed in cells exposed to the fine fraction of APS-derived particles. Therefore, our data suggest that mechanisms involved in the genotoxicity of the tested thermal-spraying-emitted particles might differ between them. While APS particles caused a marked increase in histone H2AX phosphorylation at serine-139 as an early cellular response to the induction of DNA double-strand breaks, 8-oxo-dG oxidative DNA lesions increased significantly after exposure to the HVOF particles, most likely caused by the increased intracellular ROS levels that were observed.

We have recently reported the effect of both fractions of the HVOF-derived particles studied herein in human 3D bronchial epithelial cultures (MucilAir™) under ALI conditions [22]. We found that PGFP aerosols affected cell viability at dose levels as low as 9 µg/cm², which was not seen for the aerosolised PGNP. However, exposure to PGNP (4.5 mg/cm²) caused an increase in the oxidative DNA damage of MucilAir™ cultures. In the present study, under submerged conditions, a stronger toxic response was observed, as both fractions of HVOF particles induced a pronounced decrease in cell viability and higher levels of oxidative DNA damage at lower concentrations after 24 h of exposure in A549 cells. Differences in the magnitude of the responses to the HVOF particles within the two studies may be explained by differences in the attained deposited doses as well as in the sensitivity of the cell models used. While aerosolised particles directly deposit on cell surfaces, in submerged conditions, particles in suspension may react with the culture medium and aggregate into larger particles [55,56]. On the other hand, human primary cultures under ALI conditions have been described as more resistant than the traditional 2D monoculture models since they exhibit a higher degree of complexity, with active ciliary beating and mucus production that mimic the mucociliary clearance defence system that occurs in vivo [57,58], which might have considerably attenuated the uptake and cellular effects of the HVOF-derived particles [22]. Our simplified test system does not account for the differences in pulmonary deposition of PGFP and PGNP, whereas it is well-established that the smaller sizes may penetrate deeper into the lung and more efficiently reach the more vulnerable alveoli. Statements on the actual human health risk are, therefore, not possible in the absence of actual personal exposure levels.

4. Materials and Methods
4.1. Chemicals
Dimethyl sulfoxide (DMSO), sodium hydroxide (NaOH), sodium chloride (NaCl), potassium chloride (KCl) and potassium hydroxide (KOH) were purchased from Merck KGaA (Darmstadt, Germany). Triton X-100, bovine serum albumin (BSA), low melting point (LMP)
agarose, Tris hydrochloride (Tris-HCl), silver nitrate (AgNO₃), 4-(2-hydroxyethyl)piperazine-1-ethanesulfonic acid (HEPES), methyl methanesulfonate (MMS), propidium iodide (PI), a Roche cytotoxicity detection kit (LDH) and cell proliferation reagent water-soluble tetrazolium (WST-1) were bought from Sigma-Aldrich (Madrid, Spain). Tris-base and disodium salt dihydrate (Na₂EDTA) were supplied from Merck Millipore (Madrid, Spain). Normal melting point (NMP) agarose was purchased from Bioline (London, UK). Potassium bromate (KBrO₃) and camptothecin were supplied from Alfa Aesar (Karlsruhe, Germany). Formamidopyrimidine-DNA glycosylase (FPG) was purchased from New England BioLabs (Ipswich, MA, USA). Guava ICF instrument cleaning fluid was supplied by Luminex Corporation (Austin, TX, USA), and pH 7.4 flow cytometry grade PBS from Life Technologies Corp. (Carlsbad, CA, USA). RNAse A from bovine pancreas (DNAse-free) from Applichem Panreac, eBioscience™ phospho-histone H2A.X (Ser139) Alexa Fluor® 488-conjugated monoclonal antibody (CR55T33), Invitrogen™ SYBR® Gold dye and CM-H₂DCFDA (General Oxidative Stress Indicator) were bought from Thermo Fisher Scientific (Madrid, Spain). All cell culture reagents were purchased from Gibco, Thermo Fisher Scientific (Madrid, Spain). All chemicals used were of high purity or analytical grade.

4.2. Fine and Nano-Sized Particle Suspensions and Characterisation

Incidental process-generated fine (PGFP; <2.5 µm MMAD) and nano-sized particles (PGNP; <0.200 µm MMAD) that are emitted during APS and HVOF spraying were collected directly from the inside of the spraying booths at an industrial-scale mechanical workshop in the vicinity of Barcelona, as previously described [2]. Regarding the APS spraying, the collection was performed during the injection of a Cr/Ni (50/50) and Al₂O₃ + TiO₂ feedstock blend, while for HVOF it was performed during the injection of a tungsten carbide (WC)—chromium carbon (CrC)–Ni–cobalt (Co) feedstock blend, as described in [2]. Both the PGFP and PGNP fractions were sampled directly as stock suspensions for toxicity testing, using an aerosol concentration enrichment system, as previously described [59]. The collected samples were subjected to gamma-ray irradiation to ensure the required sterility prior to the cell incubations.

The hydrodynamic size and concentration (number of particles/mL) of the aqueous particle suspensions under study were determined by a nanoparticle tracking analysis (NTA) using a NanoSight LM20 (NANOSIGHT Ltd., Salisbury, UK). The particle oxidative potential (acellular ROS production) was determined by electron spin resonance (ESR) based on the trapping of NP-induced hydroxyl radicals (OH) that are generated in the presence of hydrogen peroxide (H₂O₂) using DMPO (5,5-dimethyl-1-pyrroline-N-oxide) as a spin trap, as previously described [60]. Briefly, the particle suspensions were mixed with 0.5 M H₂O₂ and 0.05 M DMPO, followed by incubation for 15 min at 37 °C in a heated shaking water bath prior to the ESR analysis. The ESR quantification was conducted with the Analysis Software (2.0 Magnettech GmbH, Berlin, Germany) on first derivation of ESR signals of the DMPO•OH quartet as the average of total amplitudes and was expressed in arbitrary units (A.U.) per sampled volume [60].

4.3. Cell Culture

Lung adenocarcinoma epithelial A549 cells from the American Type Culture Collection (ATCC®, CCL-185™) were cultured in RPMI 1640 medium with Glutamax™ and 25 mM HEPES and supplemented with 10% heat-inactivated foetal bovine serum (FBS), 50 U/mL penicillin and 50 µg/mL streptomycin. Cells were maintained in a humidified atmosphere with 5% CO₂ at 37 °C. To carry out the submerged exposure experiments, cells were seeded in 96-well (1.0 × 10⁴ cells/well) or 24-well plates (5.0 × 10⁴ cells/well) and allowed to adhere for 48 h at 37 °C and 5% CO₂.

4.4. Exposure Conditions

All particle stock suspensions under study were dispersed by indirect probe sonication using a Branson sonifier (model 450) equipped with a disruptor cup horn accord-
ing with the standard operation procedure (SOP) for the preparation of NP suspensions developed within the NanoToxClass project [61]. The selected concentrations of each particle fraction depended on the stock’s concentrations and the volume available. In addition, the tested concentrations were chosen based on the daily alveolar mass dose of 0.13 µg/cm², which was expected to achieve a worst-case occupational exposure scenario with a maximum accumulated lifetime dose of 420 µg/cm², according to Paur et al. [62]. Accordingly, serial dilutions of the stock suspensions were carried out in incubation media (FBS-free cell culture medium), and A549 cells were exposed for 24 h at 37 °C and 5% CO₂.

For APS, the tested particle concentrations ranged from 0.31 to 5.00 µg/cm² and from 0.16 to 2.50 µg/cm² for PGFP and PGNP, respectively, while, for HVOF particles, the concentrations ranged between 2.50 and 40.00 µg/cm² and between 0.63 and 10.00 µg/cm² for PGFP and PGNP, respectively. At least three independent experiments with three replicates each were performed.

4.5. Cytotoxicity Assessment

To assess the impact of the tested particles in human alveolar epithelial cells, two endpoints were evaluated: LDH release and WST-1 reduction, indicators of plasma membrane integrity and cell metabolic activity, respectively.

LDH release was determined using a Roche Cytotoxicity Detection Kit (Roche, Mannheim, Germany) according to the manufacturer’s instructions. After exposure and prior to analysis, samples were carefully transferred to 96-well round bottom plates and then centrifuged at 2210 × g for 5 min to remove the cell debris and residual (nano)particles. Cells lysed with 2% Triton X-100 (30 min) were used as PC. Briefly, 100 µL of freshly prepared reaction mixture was added to 100 µL of each sample and incubated up to 30 min at room temperature and protected from light. The absorbance was measured at 490 nm and 630/690 nm (reference wavelength) in a microplate reader (SpectraMax® iD3 Molecular Devices, San Jose, CA, USA). The LDH release values were normalised considering the PC mean value (total LDH release).

The cell metabolic activity and viability were assessed using the WST-1 Cell Proliferation Reagent Kit (Roche, Mannheim, Germany), according to the manufacturer’s instructions. After exposure and prior to analysis, cells were washed with PBS pH 7.4. Afterwards, 100 µL/well of WST-1 reagent diluted to 1:10 was added for a 2 h incubation period at 37 °C and 5% CO₂. The sample’s absorbance was measured at 450 nm and 630/690 nm (reference wavelength) in a microplate reader (SpectraMax® iD3 Molecular Devices, San Jose, CA, USA). The WST-1 reduction values were normalised considering the NC mean value.

To test for possible particle interferences with the assays, the PC was determined in the absence and in the presence of the highest tested concentration of particle liquid suspensions. None of tested particles seemed to interfere in the conducted cytotoxicity assays since no significant differences in the PC values in the absence vs. in the presence of the highest tested concentration of PGFP and PGNP were detected.

4.6. Intracellular Reactive Oxygen Species Generation

The generation of ROS was estimated using the Invitrogen™ CM-H₂DCFDA General Oxidative Stress Indicator probe (Thermo Fisher Scientific, Madrid, Spain) according to the manufacturer’s instructions. Briefly, 1.0 × 10⁴ cells/well were seeded in 96-well black clear-bottom plates and the medium was renewed after 24 h. At 48 h post-seeding, cells were loaded with 5 µM CM-H₂DCFDA probe for 1 h at 37 °C and 5% CO₂. Then, the medium was aspirated, and cells were exposed to the tested particles over a 24 h period. Following exposure, the fluorescence was measured at 492 nm/527 nm (excitation/emission) in a microplate reader (SpectraMax® iD3 Molecular Devices, San Jose, CA, USA). The ROS production was normalised considering the mean fluorescence (arbitrary units) of the NC.
4.7. Cellular Uptake, Cell Cycle and Histone Gamma-H2AX Phosphorylation Analysis by Flow Cytometry

The determination of cellular particle uptake, changes in the cell cycle by determining cellular distribution in the different phases (G0/G1, S, G2/M and Sub-G1) and DNA double-strand breaks assessed via phosphorylation of the Ser-139 residue of the histone variant H2AX (γ-H2AX) were carried out by flow cytometry using a Guava® easyCyte™ flow cytometer (Merck KGaA, Darmstadt, Germany). Briefly, 5.0 × 10^4 cells/well were seeded onto 24-well plates, with medium renewal after 24 h. At 48 h post-seeding, cells were exposed to three non-cytotoxic concentrations of each particle over a 24 h period, followed by medium removal and the washing of the cells with pH 7.4 PBS. Cells were detached using 0.05% trypsin-EDTA, inactivated with incubation medium and centrifuged at 900 × g for 5 min. The supernatant was gently removed, and the cells were permeabilised and fixed with ice-cold 70% ethanol and left overnight at −20 °C. To remove the ethanol, the samples were centrifuged, washed with PBS with 1% BSA and once again centrifuged at 900 × g for 5 min. Then, the cells were labelled with 5 µg/mL Phospho-Histone H2A.X (Ser139) Alexa Fluor® 488-conjugated monoclonal antibody for 15 min at room temperature and protected from light, followed by a washing step with PBS with 1% BSA and centrifugation at 900 × g for 5 min. Prior to analysis, a final 15 min staining at room temperature with a 50 µg/mL RNase and 50 µg/mL PI solution was performed to ensure that only nuclear DNA was stained. Acquisitions were made with approximately 5000 events/per sample and recorded at a low flow rate (0.24 µL/s). For estimating the potential of particles to enter cells, the analysis was carried out by measuring the size (forward scatter, FSC) and complexity (side scatter, SSC) of the cells, following the protocol described by Suzuki, et al. [63]. Debris and doublets were gated out by plotting SSC-width vs SSC-area. A cell cycle analysis was performed by evaluating the relative cellular DNA content from the PI signal detection, as previously described by Rosário, et al. [64], while γ-H2AX was measured by assessing the Alexa Fluor® 488 and PI channel intensities, as described by Valdiglesias, et al. [65]. Camptothecin at 3.5 µg/mL was used as a PC to help define where cells were positive for γ-H2AX. The data were analysed using the Guava® InCyte™ Software (Merck KGaA, Darmstadt, Germany).

4.8. DNA Damage Assessment

Primary and oxidative DNA damage were assessed by the standard alkaline and FPG-modified comet assay versions, respectively, as previously described [60]. The Minimum Information for Reporting Comet Assay procedures and results (MIRCA) recommendations were followed in this manuscript [66]. Briefly, 5.0 × 10^4 cells/well were seeded onto 24-well plates, with medium renewal after 24 h. At 48 h post-seeding, cells were exposed for 24 h to three non-cytotoxic concentrations of each tested particle. After exposure, cells were washed twice with PBS pH 7.4, scrapped and suspended in PBS pH 7.4. Cells exposed to 500 µM MMS and 2.5 mM of KBrO3 for 30 min were included as a PC for primary and oxidative DNA damage assessment, respectively. The cells were counted in a Neubauer’s chamber and 6.0 × 10^3 cells were transferred to a microcentrifuge tube, centrifuged at 700 × g for 5 min and then embedded in 100 µL of 1% LMP agarose. Five µL of each sample was placed on microscope slides precoated with 1% NMP using a high-throughput system of a 12-gel comet assay unit (Severn Biotech Ltd.®, Kidderminster, UK) and placed for 5 min at 4 °C. Duplicates of each sample were added per slide. Slides were performed in triplicate, one for the alkaline version and the other two for the enzymatic version of the comet assay (with or without the FPG enzyme). Then, slides were immersed in an ice-cold lysis solution (NaCl 2.5 M, Na₂EDTA 100 mM, Tris-base 10 mM, NaOH 10 M, pH 10, Triton-X 100 1%) for 2 h at 4 °C. For the enzymatic version, the slides were washed in freshly prepared ice-cold buffer F solution (HEPES 400 mM, KCl 1 M, Na₂EDTA 5 mM, BSA 2 mg/mL, pH 8.0) (3 × 5 min) at 4 °C. A 30 µL solution of buffer F or FPG enzyme (2.7 U/mL) was added to each well and incubated for 30 min at 4 °C. For this to happen, slides were previously placed on an ice-cold metal base of a 12-gel chamber apparatus (Severn Biotech Ltd.,
Kidderminster, UK) covered by a silicon rubber 12-well mould, followed by a top plate and clamped. Meanwhile, slides for the alkaline version remained in lysis solution. All slides were immersed in electrophoresis solution (Na₂EDTA 1 mM, and NaOH 0.3 M, pH 13) in the electrophoresis platform for 40 min, followed by electrophoresis for 30 min at a constant 25 V (0.9 V/cm) and 400 mA. For slide washing, these were first covered by cold PBS (pH 7.2) and then by deionised H₂O for 10 min each. At the end of electrophoresis, slides were neutralised and fixed as described elsewhere [67]. For comet scoring, slides were initially hydrated in TE buffer (Tris-HCl 10 mM and EDTA 1 mM, pH 7.5–8) and stained at room temperature with a 1:10,000 dilution of SYBR® Gold in TE buffer for 40 min. Comets were visualised using a Motic BA410 ELITE Series microscope equipped with a Complete EPI-Fluorescence Kit and scored using the Comet Assay IV image analysis software (Perceptive Instruments, Staffordshire, UK). At least 100 cells/experimental group (50 in each replicate gel) were scored. The comet tail DNA percentage (% tail intensity) was used as a DNA damage descriptor.

4.9. Statistical Analysis

A statistical analysis was performed using SPSS (version 26.0, Armonk, NY, USA) and GraphPad Prism (version 6.0, San Diego, CA, USA) statistical software. The experimental data were expressed as means ± standard deviations (SD). The data were tested for normality and homogeneity of variances by the Shapiro–Wilk and Levene’s tests, respectively. For each assessed timepoint, the differences between the tested concentrations and NC were estimated using a one-way ANOVA followed by a post-hoc Dunnett’s test for multiple comparisons. A p value < 0.05 was considered significant.

5. Conclusions

HVOF particles were more cytotoxic compared to APS particles, most likely due to differences in their elemental composition. As hypothesised, PGNP derived from HVOF were more cytotoxic to A549 cells than PGFP, while both fractions of APS-emitted particles did not induce significant cytotoxic effects in A549 cells. Notwithstanding, particles emitted from the two thermal spraying processes under study were genotoxic to human alveolar epithelial cells. While APS particles markedly increased the levels of H2AX phosphorylation, HVOF particles caused 8-oxo-dG oxidative DNA lesions, and the effects were evident at lower concentrations of PGNP compared to PGFP.

Our data highlight that workers from industries employing high-energy processes may be at (increased) risk of adverse health effects depending on the actual inhaled dose, i.e., exposure levels and duration. Occupational epidemiological studies are urgently needed to establish this risk, whereas a better understanding of the cellular mechanisms involved in process-generated particle-induced biological effects, ultimately contributing to controlling exposures to these particles in the workplace, would facilitate reducing the health risks. The availability of information obtained from real-world exposure scenarios is deemed essential to establishing realistic preventive and corrective measures adapted to the different work scenarios (manufacturing technologies and/or chemical composition of materials).

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Appendix A

| Particle uptake | Cell cycle | γ-H2AX |
|-----------------|-----------|--------|
| Plot P01, ungated | Plot P05, gated on Pop1 R10 | Plot P04, gated on Pop1 R10 |
| Negative control | Count | Green B Fluorescence (RED-B-FLLin) |
| Side Scatter (SSC-HLIn) | 20, 40, 60, 80 | Red-B Fluorescence.e (RED-B-FLLin) |
| PFP, 5 µg/cm² | Count | Green B Fluorescence (RED-B-FLLin) |
| Side Scatter (SSC-HLIn) | 20, 40, 60, 80 | Red-B Fluorescence.e (RED-B-FLLin) |
| PGGF, 1.25 µg/cm² | Count | Green B Fluorescence (RED-B-FLLin) |
| Side Scatter (SSC-HLIn) | 20, 40, 60, 80 | Red-B Fluorescence.e (RED-B-FLLin) |

Figure A1. Representative flow cytometry graphs obtained for particle uptake, cell cycle and γ-H2AX analysis in human alveolar epithelial cells incubated for 24 h with the highest tested concentration of process-generated fine (PGFP) and nano-sized (PGNP) particles released during high-velocity oxy-fuel spraying (HVOF).
Table A1. Summary of DNA damage in cells exposed to process-generated fine (PGFP) and nano-sized (PGNP) particles released during high-velocity oxy-fuel spraying (HVOF) and respective experimental controls.

| Condition          | Primary DNA damage | Oxidative DNA damage |
|--------------------|--------------------|----------------------|
| Negative control   |                    |                      |
| Positive control   |                    |                      |
| PGFP 5 µg/cm²      |                    |                      |
| PGNP 1.25 µg/cm²   |                    |                      |

Figure A2. Representative comet assay images (100× magnification) of human alveolar epithelial cells exposed to the highest tested concentration of process-generated fine (PGFP) and nano-sized (PGNP) particles released during high-velocity oxy-fuel spraying (HVOF) and respective experimental controls.

References

1. Bessa, M.J.; Brandão, F.; Viana, M.; Gomes, J.F.; Monfort, E.; Cassée, F.R.; Fraga, S.; Teixeira, J.P. Nanoparticle exposure and hazard in the ceramic industry: An overview of potential sources, toxicity and health effects. Environ. Res. 2020, 184, 109297. [CrossRef]
2. Viana, M.; Salmatonidis, A.; Bezas, S.; Ribalta, C.; Moreno, N.; Córdoba, P.; Cassée, F.R.; Boere, J.; Fraga, S.; Teixeira, J.P.; et al. Characterizing the Chemical Profile of Incidental Ultrafine Particles for Toxicity Assessment Using an Aerosol Concentrator. Ann. Work Exp. Health 2021, 65, 966–978. [CrossRef]
3. Ribalta, C.; Koivisto, A.J.; Salmatonidis, A.; López-Lilao, A.; Monfort, E.; Viana, M. Modeling of High Nanoparticle Exposure in an Indoor Industrial Scenario with a One-Box Model. Int. J. Environ. Res. Public Health 2019, 16, 1695. [CrossRef]
4. Salmatonidis, A.; Viana, M.; Pérez, N.; Alastuey, A.; Germán, F.; Angurel, L.A.; Sanfélix, V.; Monfort, E. Nanoparticle formation and emission during laser ablation of ceramic tiles. J. Aerosol Sci. 2018, 126, 152–168. [CrossRef]
5. Salmatonidis, A.; Viana, M.; Biskos, G.; Bezas, S. Particle size distributions and hygroscopic restructuring of ultrafine particles emitted during thermal spraying. Aerosol Sci. Technol. 2020, 54, 1359–1372. [CrossRef]
6. Fonseca, A.S.; Maragkidou, A.; Viana, M.; Querol, X.; Hämmer, K.; de Francisco, I.; Estepa, C.; Borrell, C.; Lennikov, V.; de la Fuente, G.F. Process-generated nanoparticles from ceramic tile sintering: Emissions, exposure and environmental release. Sci. Total Environ. 2016, 565, 922–932. [CrossRef]
7. Fonseca, A.; Viana, M.; Querol, X.; Moreno, N.; de Francisco, I.; Estepa, C.; de La Fuente, G. Ultrafine and nanoparticle formation and emission mechanisms during laser processing of ceramic materials. J. Aerosol Sci. 2015, 88, 48–57. [CrossRef]
8. Viana, M.; Fonseca, A.S.; Querol, X.; López-Lilao, A.; Carpio, P.; Salmatonidis, A.; Monfort, E. Workplace exposure and release of ultrafine particles during atmospheric plasma spraying in the ceramic industry. Sci. Total Environ. 2017, 599–600, 2065–2073. [CrossRef]
9. Salmatonidis, A.; Ribalta, C.; Sanfélix, V.; Bezas, S.; Biskos, G.; Vulpoi, A.; Simion, S.; Monfort, E.; Viana, M. Workplace Exposure to Nanoparticles during Thermal Spraying of Ceramic Coatings. Ann. Work Exp. Health 2018, 63, 91–106. [CrossRef]
10. Fonseca, A.; Viana, M.; Querol, X.; Moreno, N.; de Francisco, I.; Estepa, C.; de la Fuente, G. Workplace Exposure to Process-Generated Ultrafine Nanoparticles in Ceramic Processes Using Laser Technology. In Indoor and Outdoor Nanoparticles; Springer: Berlin/Heidelberg, Germany, 2013; pp. 159–179.

11. Stuart, B.O. Deposition and clearance of inhaled particles. Environ. Health Perspect 1984, 55, 369–390. [CrossRef]

12. Geiser, M.; Kreyling, W.G. Deposition and biokinetics of inhaled nanoparticles. Part. Fibre Toxicol. 2010, 7, 2. [CrossRef]

13. Xing, Y.-F.; Xu, Y.-H.; Shi, M.-H.; Lian, Y.-X. The impact of PM2.5 on the human respiratory system. J. Thorac. Dis. 2016, 8, E69–E74.

14. Schraufnagel, D.E. The health effects of ultrafine particles. Exp. Mol. Med. 2020, 52, 311–317. [CrossRef]

15. Karanasiou, A.; Viana, M.; Querol, X.; Moreno, T.; de Leeuw, F. Assessment of personal exposure to particulate air pollution during commuting in European cities—recommendations and policy implications. Sci. Total Environ. 2014, 490, 785–797. [CrossRef]

16. Stone, V.; Miller, M.R.; Clift, M.J.D.; Elder, A.; Mills, N.L.; Moller, P.; Schins, R.P.F.; Vogel, U.; Kreyling, W.G.; Alstrup Jensen, K.; et al. Nanomaterials Versus Ambient Ultrafine Particles: An Opportunity to Exchange Toxicology Knowledge. Environ. Health Perspect 2017, 125, 10602. [CrossRef]

17. Anderson, J.O.; Thundiyil, J.G.; Stolbach, A. Clearing the air: A review of the effects of particulate matter air pollution on human health. J. Med. Toxicol. 2012, 8, 166–175. [CrossRef]

18. Hamanaka, R.B.; Mutlu, G.M. Particulate Matter Air Pollution: Effects on the Cardiovascular System. Front. Endocrinol. 2018, 9, 680. [CrossRef]

19. Trethewan, W.N.; Burge, P.S.; Rossiter, C.E.; Harrington, J.M.; Calvert, I.A. Study of the respiratory health of employees in seven European plants that manufacture ceramic fibres. Occup. Environ. Med. 1995, 52, 97–104. [CrossRef]

20. Jaakkola, M.S.; Sripaiboonki, P.; Jaakkola, J.J. Effects of occupational exposures and smoking on lung function in tile factory workers. Int. Arch. Occup. Environ. Health 2011, 84, 151–158. [CrossRef]

21. Kargar, F.; Shahtaheri, S.J.; Golbabaie, F.; Barkhordari, A.; Rahimi-Froushani, A.; Khadem, M. Evaluation of Occupational Exposure of Glaziers of a Ceramic Industry to Cobalt Blue Dye. Iran J. Public Health 2013, 42, 868–875.

22. Bessa, M.J.; Brandão, F.; Fokkens, P.; Cassee, F.R.; Salmatonidis, A.; Viana, M.; Vulpoi, A.; Simon, S.; Monfort, E.; Teixeira, J.P.; et al. Toxicity assessment of industrial engineered and airborne process-generated nanoparticles in a 3D human airway epithelial in vitro model. Nanotoxicology 2021, 15, 542–557. [CrossRef] [PubMed]

23. Cediel-Ulloa, A.; Isaxon, C.; Eriksson, A.; Primetzhofer, D.; Sortica, M.A.; Haag, L.; Derr, R.; Hendriks, G.; Lönahl, J.; Gudmundsson, A.; et al. Toxicity of stainless and mild steel particles generated from gas–metal arc welding in primary human small airway epithelial cells. Sci. Rep. 2021, 11, 21846. [CrossRef] [PubMed]

24. Pavlovska, I.; Ramata-Stunda, A.; Martinsone, Z.; Boroduskis, M.; Patetko, L.; Martinsone, I.; Seile, A.; Vanadzins, I. In vitro impact preliminary assessment of airborne particulate from metalworking and woodworking industries. Sci. Rep. 2021, 11, 20181. [CrossRef] [PubMed]

25. Manke, A.; Wang, L.; Rojanasakul, Y. Mechanisms of nanoparticle-induced oxidative stress and toxicity. Biomed. Res. Int. 2013, 2013, 942916. [CrossRef]

26. Peixoto, M.S.; de Oliveira Galvão, M.E.; Batistuzzo de Medeiros, S.R. Cell death pathways of particulate matter toxicity. Chemosphere 2017, 188, 32–48. [CrossRef]

27. Jederlinic, P.J.; Abraham, J.L.; Churg, A.; Himmelstein, J.S.; Epler, G.R.; Gaensler, E.A. Pulmonary fibrosis in aluminum oxide Fibroblasts. Biomed. Environ. Sci. 2010, 661–669.

28. Björ, O.; Damber, L.; Edström, C.; Nilsson, T. Long-term follow-up study of mortality and the incidence of cancer in a cohort of workers at a primary aluminium smelter in Sweden. Scand J. Work Environ. Health 2008, 34, 463–470. [CrossRef]

29. Thomassen, Y.; Koch, W.; Dunkhorst, W.; Ellingsen, D.G.; Skaugset, N.P.; Jordbekken, L.; Arne Drables, P.; Weinbruch, S. Ultrafine particles at workplaces of a primary aluminium smelter. J. Environ. Monit. 2006, 8, 127–133. [CrossRef]

30. Park, E.J.; Lee, S.J.; Lee, G.H.; Kim, D.W.; Yoon, C.; Lee, B.S.; Kim, Y.; Chang, J.; Lee, K. Comparison of subchronic immunotoxicity of four different types of aluminum-based nanoparticles. J. Appl. Toxicol. 2018, 38, 575–584. [CrossRef]

31. Salnikow, K.; Zhitkovich, A. Genetic and Epigenetic Mechanisms in Metal Carcinogenesis and Cocarcinogenesis: Nickel, Arsenic, and Chromium. Chem. Res. Toxicol. 2008, 21, 28–44. [CrossRef]

32. Halasova, E.; Adamkov, M.; Matakova, T.; Kavcova, E.; Poliacek, I.; Singljar, A. Lung cancer incidence and survival in chromium exposed individuals with respect to expression of anti-apoptotic protein survivin and tumor suppressor P53 protein. Eur. J. Med. Res. 2010, 15, 65. [CrossRef] [PubMed]

33. Lippmann, M.; Chen, L.-C. Health effects of concentrated ambient air particulate matter (CAPs) and its components. Crit. Rev. Toxicol. 2009, 39, 865–913. [CrossRef] [PubMed]

34. Phillips, J.I.; Green, F.Y.; Davies, J.C.; Murray, J. Pulmonary and systemic toxicity following exposure to nickel nanoparticles. Am. J. Ind. Med. 2010, 53, 763–767. [CrossRef]

35. Kim, Y.-S.; Chung, Y.-H.; Seo, D.-S.; Choi, H.-S.; Lim, C.-H. Twenty-Eight-Day Repeated Inhalation Toxicity Study of Aluminum Oxide Nanoparticles in Male Sprague-Dawley Rats. Toxicol. Res. 2018, 34, 343–354. [CrossRef]

36. Kim, I.-S.; Baek, M.; Choi, S.-J. Comparative cytotoxicity of Al2O3, CeO2, TiO2 and ZnO nanoparticles to human lung cells. J. Nanosci. Nanotechnol. 2010, 10, 3453–3458. [CrossRef]

37. Zhang, X.Q.; Yin, L.H.; Tang, M.; Pu, Y.P. ZnO, TiO2, SiO2, and Al2O3 Nanoparticles-induced Toxic Effects on Human Fetal Lung Fibroblasts. Biomed. Environ. Sci. 2011, 24, 661–669.
38. Ivask, A.; Titma, T.; Visnapuu, M.; Vija, H.; Kakinen, A.; Sihtmae, M.; Pokhrel, S.; Madler, L.; Heinlaan, M.; Kisdan, V. Toxicity of 11 metal oxide nanoparticles to three mammalian cell types in vitro. Curr. Top. Med. Chem. 2015, 15, 1914–1929. [CrossRef]

39. Tsai, A.; Jones, E.; Case, C.P. The in vitro genotoxicity of orthopaedic ceramic (Al2O3) and metal (CoCr alloy) particles. Mutat. Res. Genet. Toxicol. Environ. Mutagenesis 2010, 697, 1–9. [CrossRef]

40. IARC. Chromium, nickel and welding. IARC Monogr. Eval. Carcinog. Risks Hum. 1990, 49, 677.

41. Wise, J.P.; Wise, S.S.; Little, J.E. The cytotoxicity and genotoxicity of particulate and soluble hexavalent chromium in human lung cells. Mutat. Res. Genet. Toxicol. Environ. Mutagenesis 2002, 517, 221–229. [CrossRef]

42. Zhang, Q.; Kusaka, Y.; Zhu, X.; Sato, K.; Mo, Y.; Kluz, T.; Donaldson, K. Comparative toxicity of standard nickel and ultrafine nickel in lung after intratracheal instillation. J. Occup. Health 2003, 45, 23–30. [CrossRef] [PubMed]

43. Morimoto, Y.; Hirohashi, M.; Ogami, A.; Oyabu, T.; Myojo, T.; Hashiba, M.; Mizuguchi, Y.; Kambara, T.; Lee, B.W.; Kuroda, E. Pulmonary toxicity following an intratracheal instillation of nickel oxide nanoparticle agglomerates. J. Occup. Health 2011, 53, 293–295. [CrossRef] [PubMed]

44. Åkerlund, E.; Cappellini, F.; Di Buchianico, S.; Islam, S.; Skoglund, S.; Derr, R.; Odneval Wallinder, I.; Hendriks, G.; Karlsson, H.L. Genotoxic and mutagenic properties of Ni and NiO nanoparticles investigated by comet assay, γ-H2AX staining, Hprt mutation assay and ToxTracker reporter cell lines. Environ. Mol. Mutagen 2018, 59, 211–222. [CrossRef] [PubMed]

45. Roedel, E.Q.; Cafasso, D.E.; Lee, K.W.M.; Pierce, L.M. Pulmonary toxicity after exposure to military-relevant heavy metal tungsten alloy particles. Toxicol. Appl. Pharmacol. 2012, 259, 74–86. [CrossRef] [PubMed]

46. Reynolds, M.; Armknecht, S.; Johnston, T.; Zhiltkovich, A. Undetectable role of oxidative DNA damage in cell cycle, cytotoxic and clastogenic effects of Cr(VI) in human lung cells with restored ascorbate levels. Mutagenesis 2012, 27, 437–443. [CrossRef]

47. DeLoughery, Z.; Luizak, M.W.; Ortega-Arienza, S.; Zhiltkovich, A. DNA Double-Strand Breaks by Cr(VI) Are Targeted to Euchromatin and Cause ATR-Dependent Phosphorylation of Histone H2AX and Its Ubiquitination. Toxicol. Sci. 2014, 143, 54–63. [CrossRef]

48. Zhang, Z.; Leonard, S.S.; Wang, S.; Vallyathan, V.; Castranova, V.; Shi, X. Cr (VI) induces cell growth arrest through hydrogen peroxide-mediated reactions. Mol. Cell. Biochem. 2001, 222, 77–83. [CrossRef]

49. Ding, J.; He, G.; Gong, W.; Wen, W.; Sun, W.; Ning, B.; Huang, S.; Wu, K.; Huang, C.; Wu, M.; et al. Effects of nickel on cyclin expression, cell cycle progression and cell proliferation in human pulmonary cells. Cancer Epidemiol. Biomark. Prev. 2009, 18, 1720–1729. [CrossRef]

50. Mo, Y.; Zhang, Y.; Zhang, Y.; Yuan, J.; Mo, L.; Zhang, Q. Nickel nanoparticle-induced cell transformation: Involvement of DNA damage and DNA repair defect through HIF-1α/miR-210/Rad52 pathway. J. Nanobiotechnol. 2018, 17, 3–79. [CrossRef]

51. Ma, A.; Dai, X. The relationship between DNA single-stranded damage response and double-stranded damage response. Cell Cycle 2018, 17, 73–79. [CrossRef]

52. Barnum, K.J.; O’Connell, M.J. Cell cycle regulation by checkpoints. Methods Mol. Biol. 2014, 1170, 29–40. [PubMed]

53. Mjelle, R.; Hegre, S.A.; Aas, P.A.; Slupphaug, G.; Drabløs, F.; Saetrom, P.; Krokam, H.E. Cell cycle regulation of human DNA repair and chromatin remodeling genes. DNA Repair 2015, 30, 53–67. [CrossRef] [PubMed]

54. Podhorecka, M.; Skladanowski, A.; Bozko, P. H2AX Phosphorylation: Its Role in DNA Damage Response and Cancer Therapy. J. Nanosci. Nanotechnol. 2010, 10, 920161. [CrossRef] [PubMed]

55. Lacroix, G.; Koch, W.; Ritter, D.; Gutleb, A.C.; Larsen, S.T.; Loret, T.; Zanetti, F.; Constant, S.; Chortarea, S.; Rothen-Rutishauser, B.; et al. Air–Liquid Interface In vitro Models for Respiratory Toxicology Research: Consensus Workshop and Recommendations. Appl. Vitr. Toxicol. 2018, 4, 91–106. [PubMed]

56. Loret, T.; Peyret, E.; Dubreuil, M.; Aguerre-Chariol, O.; Bressot, C.; le Bihan, O.; Amodeo, T.; Trouiller, B.; Braun, A.; Egles, C.; et al. Air–liquid interface to aerosols of poorly soluble nanomaterials induces different biological activation levels compared to exposure to suspensions. Part. Fibre Toxicol. 2016, 13, 58. [CrossRef] [PubMed]

57. Frieke Kuper, C.; Gröllers-Mulderij, M.; Maarschalkerweerd, T.; Meulendijks, N.M.M.; Reus, A.; van Acker, F.; Zondervan-van den Berg, H.L. Genotoxic and mutagenic properties of Ni and NiO nanoparticles investigated by comet assay, γ-H2AX staining, Hprt mutation assay and ToxTracker reporter cell lines. Environ. Mol. Mutagen 2018, 59, 211–222. [CrossRef] [PubMed]

58. George, I.; Uboldi, C.; Bernard, E.; Sobrido, M.S.; Dine, S.; Hagège, A.; Vrel, D.; Herlin, N.; Rose, J.; Orsière, T.; et al. Toxicological Assessment of ITER-Like Tungsten Nanoparticles Using an In vitro Human Airway Epithelium Model. Nanomaterials 2019, 9, 1374. [CrossRef]

59. Kim, S.; Jaques, P.A.; Chang, M.; Friones, J.R.; Sioutas, C. Versatile aerosol concentration enrichment system (VACES) for simultaneous in vivo and in vitro evaluation of toxic effects of ultrafine, fine and coarse ambient particles Part I: Development and laboratory characterization. J. Aerosol Sci. 2001, 32, 1281–1297. [CrossRef]

60. Bessa, M.J.; Brandão, F.; Fokkens, P.H.B.; Leseman, D.; Boere, A.J.F.; Cassee, F.R.; Salamontidis, A.; Viana, M.; Vulpoi, A.; Simon, S.; et al. In vitro Toxicity of Industrially Relevant Engineered Nanoparticles in Human Alveolar Epithelial Cells: Air-Liquid Interface versus Submerged Cultures. Nanomaterials 2021, 11, 3225. [CrossRef]

61. NanoToxclass Standard Operation Procedure—Preparation of Nanoparticle Suspensions by Cup Horn Sonication. Available online: https://www.nanopartikel.info/files/projekte/NanoToxClass/NanoToxClassSOP_Dispersion_by_cup_horn_sonication_V2.0.pdf (accessed on 1 December 2021).
62. Paur, H.-R.; Cassee, F.R.; Teeguarden, J.; Fissan, H.; Diabate, S.; Auferheide, M.; Kreyling, W.G.; Hänninen, O.; Kasper, G.; Riediker, M. In-vitro cell exposure studies for the assessment of nanoparticle toxicity in the lung—A dialog between aerosol science and biology. *J. Aerosol Sci.* 2011, 42, 668–692. [CrossRef]

63. Suzuki, H.; Toyooka, T.; Ibuki, Y. Simple and easy method to evaluate uptake potential of nanoparticles in mammalian cells using a flow cytometric light scatter analysis. *Environ. Sci. Technol.* 2007, 41, 3018–3024. [CrossRef] [PubMed]

64. Rosário, F.; Bessa, M.J.; Brandão, F.; Costa, C.; Lopes, C.B.; Estrada, A.C.; Tavares, D.S.; Teixeira, J.P.; Reis, A.T. Unravelling the Potential Cytotoxic Effects of Metal Oxide Nanoparticles and Metal(Loid) Mixtures on A549 Human Cell Line. *Nanomaterials* 2020, 10, 447. [CrossRef] [PubMed]

65. Valdiglesias, V.; Laffon, B.; Fásaro, E.; Méndez, J. Evaluation of okadaic acid-induced genotoxicity in human cells using the micronucleus test and γH2AX analysis. *J. Toxicol. Environ. Health A* 2011, 74, 980–992. [CrossRef] [PubMed]

66. Møller, P.; Azqueta, A.; Boutet-Robinet, E.; Koppen, G.; Bonassi, S.; Milić, M.; Gajski, G.; Costa, S.; Teixeira, J.P.; Costa Pereira, C.; et al. Minimum Information for Reporting on the Comet Assay (MIRCA): Recommendations for describing comet assay procedures and results. *Nat. Protoc.* 2020, 15, 3817–3826. [CrossRef] [PubMed]

67. Bessa, M.J.; Brandão, F.; Querido, M.M.; Costa, C.; Pereira, C.C.; Valdiglesias, V.; Laffon, B.; Carriere, M.; Teixeira, J.P.; Fraga, S. Optimization of the harvesting and freezing conditions of human cell lines for DNA damage analysis by the alkaline comet assay. *Mutat. Res.* 2019, 845, 402994. [CrossRef] [PubMed]