EXAFS and DFT study of the cadmium and lead adsorption on modified silica nanoparticles

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\textbf{Highlights}

- Modified silica nanoparticles were prepared and characterized by FTIR and \textsuperscript{13}C NMR.
- Isotherms for the adsorption of \textsuperscript{2}Cd\textsuperscript{2+} and \textsuperscript{2}Pb\textsuperscript{2+} ions on the particles were obtained.
- Pb–N and Pb–C distances on the particles were obtained by EXAFS.
- DFT calculations supported the experimental data.
- Hybrid nanoparticles by precipitation allowed the easy removal of the nano-sorbents.

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\textbf{Abstract}

Silica nanoparticles of 7 nm diameter were modified with (3-aminopropyl) triethoxysilane (APTES) and characterized by CP-MAS \textsuperscript{13}C and \textsuperscript{29}Si NMR, FTIR, zeta potential measurements, and thermogravimetry. The particles were shown to sorb successfully divalent lead and cadmium ions from aqueous solution. Lead complexation with these silica nanoparticles was clearly confirmed by EXAFS (Extended X-ray Absorption Fine Structure) with synchrotron light measurements. Predicted Pb–N and Pb–C distances obtained from quantum-chemical calculations are in very good agreement with the EXAFS determinations. The calculations also support the higher APTES affinity for Pb\textsuperscript{2+} compared to Cd\textsuperscript{2+}.

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1. Introduction

Effluents from industrial applications including mining, refining and production of textiles, paints and dyes may contain heavy metal ions at high concentrations [1]. In particular, lead and...
cadmium are considered priority metals in the framework of the European water policy (Directive 2013/39/EU [2]), because they present a significant risk to biota and humans, given its persistence, toxicity and bioaccumulation characteristics.

A wide variety of techniques to remove heavy metals from water is available, such as ion exchange, reverse osmosis and nanofiltration, precipitation, coagulation/co-precipitation and sorption [3]. Among them, adsorption is considered as one of the most suitable methods due to its ease of operation and the availability of a wide range of adsorbents. Organofunctionalization of adsorbents for the removal of heavy metal ions from water has attracted great research interest in the last decade due to the advantages of achieving high efficiency and good selectivity [4,5]. In particular, amines have been used to functionalize silica [5] to synthesize sorbents for heavy metal ions removal from aqueous media. Mesoporous silica materials chemically modified with aminopropyl, [2-aminoethylamino]-propyl, and [(2-aminoethylamino)-ethylamino]-propyl [6] were recently employed to sorb divalent copper, nickel, lead, cadmium and zinc from aqueous solution.

In this context we here prepared 7 nm-diameter silica nanoparticles organofunctionalized with aminopropyl groups to sorb divalent lead and cadmium ions from aqueous solution. Sorption isotherms for these cations on the silica particles were favorably obtained. Complexation of Pb\(^{2+}\) with the silica nanoparticles was confirmed by EXAFS (Extended X-ray Absorption Fine Structure) with synchrotron light measurements. To compare predicted Pb–N and Pb–C distances in the modified nanoparticles bonded to Pb\(^{2+}\) with the EXAFS determinations, quantum-chemical calculations were carried out.

Separation of small silica nanoparticles by filtration or ultracentrifugation can be a cumbersome procedure. Thus, to facilitate the separation of the nanomaterial by filtration, after sorption of the metal ions an alternative novel method involving the synthesis of bigger composite particles (NPNH2@MEH-PPV) of the copolymer poly[2-methoxy-5-(2-ethylhexyl-oxy)-1,4-phenylene-vinylene] (MEH-PPV) from Sigma–Aldrich were used without further purification. Distilled water (>18 MΩ cm\(^{-1}\), <20 µg L\(^{-1}\) organic carbon) was obtained from a Millipore (Bedford, MA) system.

2. Materials and methods

2.1. Materials

Fumed silica (Sigma, specific surface area; SSA = (390 ± 40) m\(^2\) g\(^{-1}\), particle diameter estimated from the SSA = 7 nm and confirmed by TEM images [7]), toluene (Baker, p.a.), ethyl acetate (Ciccarelli, p.a.), CaH\(_2\) (Fluka), Pb(NO\(_3\)\(_2\)) (Timper), ethyldrofuran (THF) Biopack, and CdSO\(_4\)·8/H\(_2\)O, CaCl\(_2\), (3-aminopropyl) triethoxysilane (APTES) and poly[2-methoxy-y-5-(2-ethylhexyl-oxy)-1,4-phenylene-vinylene] (MEH-PPV) from Sigma–Aldrich were used without further purification. Distilled water (>18 MΩ cm\(^{-1}\), <20 µg L\(^{-1}\) organic carbon) was obtained from a Millipore (Bedford, MA) system.

2.2. Modification and characterization of silica nanoparticles

Commercial fumed silica was first dried for 15 h at 150 °C. Activation of silanol groups was subsequently achieved by heating at 250 °C for 3 h. Modified silica nanoparticles were synthesized by a variation of the method published by Foscheira et al. [6]. Briefly, 1.0 g of previously dried and activated silica was mixed with 80 mL of toluene and 1.5 mL of (3-aminopropyl) triethoxysilane. The mixture was heated and maintained under reflux (110–120 °C) for 48 h. To preserve a dry environment a guard tube filled with calcium chloride was connected to the reflux condenser. Modified nanoparticles (from now named NPNH2) were recovered by evaporating the reaction solvent (toluene), washing gently with ethyl acetate, filtering through 0.22 µm nylon membrane (Osmonics Inc.) and drying at 0.1 Torr for 5 h.

The conditions of the CP-MAS \(^{13}\)C and \(^{29}\)Si nuclear magnetic resonance (NMR) experiments were described elsewhere [7].

The zeta potential measurements (0.5 g L\(^{-1}\) water suspension at constant ionic strength of 10 \(^{-3}\) M KCl) were performed at 25.0 ± 0.1 °C with a Brookhaven 90Plus/Bi-MAS instrument, operating at \(\lambda = 635\) nm from a 15 mW-solid state laser.

Fig. 1. (a) Molecular formula of MEH-PPV. (b) Synthesis of NPNH\(_2\)@MEH-PPV. Inset: TEM images of the hybrid NPNH\(_2\)@MEH-PPV nanoparticles.
For the FTIR measurements a Varian 660 spectrometer equipped with a DTGS detector was employed. The spectra were acquired in the transmission mode with a spectral resolution of 4 cm$^{-1}$ accumulating 128 scans.

The thermogravimetric curve was performed with a DuPont Model 951 instrument, under Ar atmosphere with heating rate of 0.17 K s$^{-1}$. The instrument was calibrated with CaC$_2$O$_4$·3H$_2$O.

Nitrogen analysis was performed by catalytic combustion with a Thermo Finnigan, CE Flash EA 1112 elemental analyzer.

The specific surface area (SSA) was evaluated by means of the Brunauer–Emmett–Teller (BET) equation applied to nitrogen adsorption isotherms as reported elsewhere [8].

2.3. Batch sorptions

A series of plastic vials containing solution of 1.0 g L$^{-1}$ of NPNH$_2$ were incubated overnight at 25.0 ± 0.1 °C with aqueous solutions of different concentration of Pb$^{2+}$ (pH = 3) or Cd$^{2+}$ (pH = 6) in a reciprocal shaker at 150 rpm. The content of each vial was then filtered through a 0.1 μm membrane. The final metal concentration was determined in the filtered solution by atomic absorption spectrophotometry (Shimadzu AA6650, Shimadzu Corporation Kyoto, Japan).

The metal uptake capacity, $Q$, was calculated from the difference between the initial ($C_i$) and equilibrium ($C_{eq}$) concentrations as shown in Eq. (1).

$$Q = \frac{(C_i - C_{eq})V}{M}$$

where, $Q$ (mmol/g) is the sorbed metal quantity per gram of sorbent at any time, $M$ (g) is the sorbent dosage, and $V$ (L) is the solution volume.

2.4. Synthesis of the hybrid nanoparticles NPNH$_2$@MEH-PPV

Multiphase particles of the copolymer MEH-PPV (Fig. 1a) and the modified silica nanoparticles containing the metal ions were prepared by a controlled precipitation method [9,10]. The procedure is based on a controlled precipitation of the polymer in aqueous medium (see Fig. 1b). Briefly, a 0.5 g L$^{-1}$ stock solution of the polymer was prepared in tetrahydrofuran (THF). A volume of 0.5 mL of the stock solution was added to 10 mL of aqueous solutions containing 0.25 mg of NPNH$_2$ and different concentration of Pb$^{2+}$ (pH = 3) or Cd$^{2+}$ (pH = 6), preincubated for 4 h at 25.0 °C in a reciprocal shaker at 150 rpm. The content of each vial was then filtered through a 0.1 μm membrane. Final metal concentration was determined in the filtrate solution by atomic absorption. Particle formation occurs by precipitation of the polymer upon rapidly adding the polymer solution to an excess of the aqueous phase [9,10].

2.5. XAFS measurements

Lead L$_3$-edges XAFS spectra were measured at the XAFS2 beam line at the Laboratorio Nacional de Luz Síncrono (LNLS, Campinas, Brazil) at room temperature in transmission mode with three ion chambers as detectors. The third one was used to measure the metallic reference simultaneously with the sample. Powder sample was pressed in a pellet optimizing the absorption step and total absorption. The EXAFS (Extended X-ray Absorption Fine Structure) oscillations $\chi(k)$ was extracted from the experimental data with standard procedures using the Athena program, part of the IFEFFIT package [11]. The Fourier transformation was calculated using the Hanning filtering function. The $k^2$-weighted $\chi(k)$ data, to enhance the oscillations at higher $k$, were Fourier transformed. EXAFS modeling was carried out using the ARTEMIS program which is also part of the IFEFFIT package. Structural parameters (coordination numbers and bond lengths and their mean squared disorders) were obtained by a nonlinear least-squares fit of the theoretical EXAFS signal to the data in $R$ space by Fourier transforming both the theory and the data. Theoretical scattering path amplitudes and phase shifts for all paths used in the fits were calculated using the FEFF6 code [12]. The $k$-range was set from 2.5 to 12 Å$^{-1}$, the Fourier transforms were fitted in region 1.2–3.2 Å and the fits were performed simultaneously to $k$, $k^2$ and $k^3$-weighted data. The passive reduction factor $S_0$ was restrained to values of 0.7. This value was obtained from fitting of metallic Pb foil standard by constraining the coordination number in this compound of known crystal structure.

2.6. Density functional theory calculations

To compare predicted Pb–N and Pb–C distances in the APTES-functionalized nanoparticles bonded to Pb$^{2+}$ with the EXAFS determinations, quantum-chemical calculations were carried out. The popular density functional hybrid method B3LYP [13–15] and the hybrid meta-generalized exchange–correlation functional M06-2X [16] combined with the CEP-121G basis set [17,18] were employed. In the CEP-121G, compact effective potentials (CEP) replace the atomic core electrons, while the valence pseudo-orbitals are described by Gaussian expansions with a triple-zeta contraction scheme (triple-split valence basis set). In particular for APTES–Pb$^{2+}$ calculations, the 44 alpha electrons and the 44 beta electrons are distributed in molecular orbitals which comprise 245 basis functions based on 359 primitive Gaussians. In comparative calculations for PTES-Cd$^{2+}$ the 52 alpha electrons and the 52 beta electrons are distributed in molecular orbitals which comprise 271 basis functions based on 401 primitive Gaussians. Bulk hydration effects were considered using the conductor-like polarizable continuum model, CPCM [19]. Molecular geometry optimizations without symmetry constraints were performed using analytical gradient methods. To derive thermodynamic properties, the harmonic vibrational frequencies were then derived via analytical second derivative methods. All calculations were performed with the Gaussian 09 program package [20].

3. Results and discussion

3.1. Characterization of NPNH$_2$

3.1.1. Infrared spectroscopy

The FTIR spectrum of NPNH$_2$ shows additional peaks compared to those reported for bare silica nanoparticles [7,21] (see Fig. 2 and Table 1).
The decrease in specific surface area upon structural units, also referred as $Q_{29}$Si NMR spectrum of the silica nanoparticles organofunctionalized with aminopropyl groups was already reported for different systems [29]. Furthermore, another two peaks at −67 and −58 ppm are observable due to the incorporation of aminopropyl groups on the surface. Also, no signal is noticeable around −45 ppm, which would correspond to the chemical shift of silicon atoms in liquid (3-aminopropyl) trialkoxysilane, indicating the absence of free silane molecules physically sorbed on the silica surface.

The grafting of organic species onto the silica surface leads to the appearance of peaks at −67 and −58 ppm in the spectrum, indicating the generation of new siloxane linkages (Si–O–Si) of the aminopropylsilane silicon to the surface silicon atoms of the silica nanoparticles. The signal at −67 ppm is attributed to the aminopropylsilane silicon attached to the surface of the material via three siloxane bonds, (–O–)3Si–R (T1; where R = aminopropyl chain), while the peak at −58 ppm is attached via two siloxane bonds, (–O–)2Si(OH)R (T2). The relative high intensity of T2 and T3 species and the absence of T1 species (–O–)Si(OH)2R indicate that the incorporated aminopropyl groups are closely packed on the surface of the material through tridentate and bidentate linkages instead of a monodentate one [29,30].

It can be seen from the dependence of the zeta potential of the NPNH2 dispersion on pH (Fig. 4) that the pH of the zero point of charge (PZC), i.e., the pH where the zeta potential equals zero, is 9.15. At pH below 9.15, the particles charge is positive due to the protonation of the amino-modification compensating the negative charge of the silanol groups. These groups are responsible for the PZC of bare silica, which is between pH 1 and 2, depending on the particle size [31,32]. For the NPNH2 in suspensions of pH above 9.15, where several of the amino groups are not protonated (pKa of

Table 1. These bands are also present in the spectrum of APTES. The bands in the 3000–2800 cm$^{-1}$ region are assigned to the CH$_2$ stretching [22]. Absorption bands in the region 1575–1450 cm$^{-1}$ may be assigned to N–H vibrations in protonated amines [23]. The N–H stretch absorption over the 3300–3400 cm$^{-1}$ is masked by the O–H absorption present in silica [21]. Any absorption due to the vibration modes assigned to the Si–O–C moiety is shrouded by the strong Si–O–Si absorption signals at 1100 cm$^{-1}$ [22]. Absorption at 1332 and 1630 cm$^{-1}$ is assigned to CO$_2$ trapped by surface NH$_3$ giving rise to an amine bicarbonate moiety in coincidence with that reported for protonated (3-aminopropyl) triethoxysilane films and aged silicon particles modified with propylamine groups [25].

The percentage of organic groups (%OG = 9.8%) for NPNH2 was estimated from the loss of mass observed from 200 to 700 °C in the TG analysis (see TG curve in Fig. S1 of the Electronic Supplementary material). This value is in excellent agreement with the 1.04 ± 0.02% N content determined by nitrogen analysis. The specific surface area (SSA) of the bare SiO$_2$ was estimated from the loss of mass observed from 200 to 700 °C, via three siloxane bonds, (–O–)$_3$Si–R (T1; where R = aminopropyl group), while the peak at −58 ppm is attributed to the aminopropylsilane silicon attached to the surface of the material via three siloxane bonds, (–O–)$_3$Si(OH)R (T2). The relative high intensity of T2 and T3 species and the absence of T1 species (–O–)Si(OH)$_2$R indicate that the incorporated aminopropyl groups are closely packed on the surface of the material through tridentate and bidentate linkages instead of a monodentate one [29,30].

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free propylamine = 10.566 [33]), the particles possess negative zeta potentials. The values of zeta potential obtained at pH below 4 are underestimated because the ionic strength cannot be considered constant in the medium containing 0.001 M KCl.

3.2. Batch sorption experiments

The amount of sorbed metal $Q$ from 100 mg L$^{-1}$ solutions of Cd$^{2+}$ or Pb$^{2+}$ reached constant values after 4 h of incubation. Thus, to assure equilibrium conditions in the batch sorption experiments the systems were incubated overnight. The percentage of sorbed ions depends on the metal and on the solution pH (see Table S1 of the Electronic Supplementary material). As can be seen in the table, the most favorable Cd$^{2+}$ sorption is achieved at pH = 6. Thus, the isotherm of Cd$^{2+}$ sorption was measured at this pH.

At very low pH, cadmium is present in solution as free Cd$^{2+}$ ion, but there is also a high concentration of H$^+$ ions that protonate the surface of the material causing a strong positive charge, which blocks the sorption of cadmium ion by electrostatic repulsion. From moderately acidic to neutral conditions, cadmium still exists mostly as free ion, but the competition effect with H$^+$ markedly decreases.

The effect of the solutions pH on the removal efficiencies of lead can also be explained by the metal speciation at different pH values [34] and the variation in the surface charge of the nanomaterial with pH. The pH for performing the batch sorption experiments with Pb$^{2+}$ was 3. At this pH ($Q = 0.195$ mmol Pb$^{2+}$/g NPNH2), lower pH values were avoided because of the competition effect with H$^+$, whereas at higher pH precipitation of the hydroxide from the solutions of higher metal concentration takes place (Ksp Pb(OH)$_2$ = 1.43$ \times 10^{-20}$) [35]. Higher pH values were also avoided because at Pb$^{2+}$ initial concentrations of 100 ppm there is precipitation at pH $\geqslant$ 5 (see Table S1 of the Electronic Supplementary material), and even higher concentrations of Pb$^{2+}$ were employed for obtaining the isotherm.

To further characterize the Pb$^{2+}$ (pH = 3) and Cd$^{2+}$ (pH = 6) sorptions on NPNH2, the isotherms shown in Fig. 5(a and b) were obtained by incubation of 1.0 g L$^{-1}$ of the nanoparticles at 25.0 °C in solutions containing different amounts of the cations.

As can be seen in Fig. 5(a), the values of Q (mmol Pb$^{2+}$/g NPNH2) linearly increase with the equilibrium concentration ($C_{eq}$) of Pb$^{2+}$ expressed in mmol L$^{-1}$ (Eq. (2)).

$$Q = (-8.2 \pm 8.7) \times 10^{-3} + (0.22 \pm 0.01) \times C_{eq}$$

where $Q$ and $C_{eq}$ are expressed in mmol Pb$^{2+}$/g NPNH2 and mmol Pb$^{2+}$/L$^{-1}$, respectively.

The Pb$^{2+}$ sorption data can also be fitted to a Langmuir isotherm (Eq. (4)).

$$Q = \frac{0.13 \pm 0.01}{1 + (1.4 \pm 0.3) \times C_{eq}}$$

where $Q$ and $C_{eq}$ are expressed in mmol Pb$^{2+}$/g NPNH2 and mmol Pb$^{2+}$/L$^{-1}$, respectively.

Langmuir sorption isotherms are characterized by saturation at high concentrations. From the parameters of Eq. (4), it can be estimated a quantitative value of monolayer for the Langmuir model, which represents the maximum sorption that corresponds to a monolayer formation of 0.093 mmol of sorbed Pb$^{2+}$ per gram of nanoparticles. This value is lower than the 0.44 mmole of the organic groups in 1 g NPNH2 calculated from the measured %OG, which indicates that not all the amine groups are available for adsorption due to aggregation of the particles.

Comparison of the metals uptake capacity shows that NPNH2 are much better sorbents of Pb$^{2+}$ than of Cd$^{2+}$. This result is in line with reported data by Aguado et al. for different silica amino-functionalized SBA-15 materials employed to remove heavy metals from water [5] and is also supported by our DFT calculations (see below).

3.3. EXAFS Analysis of the Nano-sorbents

After Pb$^{2+}$ sorption experiments, the suspensions were filtered with 0.1 µm nylon filters and the nanoparticles were gently washed with water, in order to confirm the complexation of the cation with the silica nanoparticles. Separation of 7 nm diameter particles with this filter is possible due to aggregation of the nanoparticles during the filtering process as already observed for other particles prepared by modification of 7 nm silica particles [7]. EXAFS experiments were performed at the Pb L$_3$ edge, to provide information of the atomic structure around the sorbent atoms. Thus, it is possible to obtain information on type, number and distances between Pb$^{2+}$ ions and their neighbors and reconstruct the atomic moiety around them. From the X-ray fluorescence spectra it was not possible to detect the presence of lead in bare silica nanoparticles previously incubated with these metal ions, which indicates that upon washing the samples with water, complete removal of the metal occurred. This indicates that lead sorption by silanol groups (if any) should be of physical nature.

The EXAFS oscillation and the corresponding Fourier Transform as well as the fitting function obtained from the NPNH2 sample are shown in Fig. 6. Three shells around Pb$^{2+}$ ions were proposed in the fit and the results are shown in Table 2. Distances longer than 3.5 Å could not be fitted accurately, probably because of the superposition of single and multiple scattering paths. The first fitted shell corresponds to one N atom at 2.28 Å, and the other two correspond to two C atoms at 2.73 and 3.38 Å, respectively. This is compatible
with the chemical sorption of Pb$^{2+}$ on N sites of the aminopropyl groups. The second and third neighbors of Pb atoms correspond to the first two C atoms of the chain. Further neighbors observed in the Fourier Transform after 3.2 Å can be associated to single and multiple scattering of the rest of the carbons chain causing a focusing effect because of the linear atoms alignment [37].

3.4. DFT calculations

To compare predicted Pb–N and Pb–C distances in the modified nanoparticles bonded to Pb$^{2+}$ with the EXAFS determinations, DFT calculations were performed. For simplicity, we assumed that the above mentioned interatomic distances in the APTES–Pb$^{2+}$ moiety are only slightly affected by the nanoparticle presence. Therefore, the single APTES–Pb$^{2+}$ molecule has been modeled. The obtained fully optimized structures and the derived relevant data are shown in Fig. 7.

The theoretical Pb–N bond distance of 2.33 Å resulting for the configuration (a) compares well with the experimental one of 2.28 Å, while the predictions for the Pb–C non-bonded distances are in reasonable agreement with the experimental values listed in Table 2. In addition, the calculations indicate that the larger Pb–C distance corresponds to the C atom linked to the N atom.

The cyclic-conformer (b) is much more stable than the conformer (a). In fact, the calculated CPCM-B3LYP/CEP-121G enthalpy change at 298 K for the APTES + Pb$^{2+}$ → APTES–Pb$^{2+}$ association process is $-51.3 \text{ kcal mol}^{-1}$ for conformer (a) and $-80.7 \text{ kcal mol}^{-1}$ for conformer (b). On the other hand, enthalpy changes of $-45.7$ and $-78.5 \text{ kcal mol}^{-1}$ were estimated for these processes at the CPCM-M06-2X/CEP-121G level of theory. Although the Pb–N and Pb–C interatomic distances in the conformer (b) (Fig. 7(b)) are also consistent with the experimental results, due to the important rearrangement involved (the minimum Pb–O distance in (a) is 5.3 Å, whereas in the cyclic conformer (b) it is only 2.27 Å), it is expected that the cyclization reaction (a) → (b) or the direct insertion reaction of the Pb$^{2+}$ into APTES involves large activation energy values. This assumption is quite consistent with the absence of the EXAFS signal corresponding to the strong Pb–O bond at 2.27 Å.

Similar considerations are expected for the APTES–Cd$^{2+}$ complexes. The corresponding open and cyclic molecular configurations are depicted in Fig. S2 of the Electronic Supplementary material. However, the second conformation is only
particles of the copolymer poly[2-methoxy-5-(2-ethylhexyl-ox
y)-1,4-phenylene-vinylene] and the modified silica nanoparticles
was also proposed.

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Appendix A. Supplementary data

Supplementary data associated with this article can be found in
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Fig. 7. Optimized structures of APTES–Pb2+. (a) open-conformer and (b) cyclic-
conformer, at the CPCM-B3LYP/CEP-121G level of theory. Interatomic distances in
Å. Experimental distances obtained by EXAFS are shown in parentheses.

−3.0 kcal mol⁻¹ (B3LYP) or −4.6 kcal mol⁻¹ (M06-2X) more stable
than the first ones. Besides, the enthalpy changes for the formation
of the open conformer of APTES–Cd²⁺ of −21.6 kcal mol⁻¹ (B3LYP)
and −22.3 kcal mol⁻¹ (M06-2X) are smaller than the obtained for
the open configuration of APTES–Pb²⁺. These results are in qualita-
tive agreement with the observed much higher sorbent efficiency of
APTES for Pb²⁺ than for Cd²⁺.

3.5. Hybrid NPNH2@MEH-PPV nanoparticles

TEM images for NPNH2@MEH-PPV hybrid particles obtained
are shown in Fig. 1(b) (inset). As observed in Fig. 1(b) rapid injec-
tion of a THF solution of MEH-PPV dissolved into the aqueous sus-
pension of NPNH2 gave average diameter particles of 300 nm. Similar
results were obtained with the suspension of NPNH2 incubated in the
Cd²⁺ solution. The procedure can be applied to the separa-
tion and removal of nanoparticles employed for the sorption of
metal ions, as exemplified here.

4. Conclusion

Silica nanoparticles of 7 nm diameter were modified with
(3-aminopropyl) triethoxysilane (APTES) and characterized by
CP-MAS ¹³C and ²⁹Si NMR, FTIR, zeta potential measurements,
and thermogravimetry. Isotherms for the adsorption of divalent
lead and cadmium ions from aqueous solution were obtained.
Lead complexation with these silica nanoparticles was clearly con-
firmed by EXAFS. Predicted Pb–N and Pb–C distances obtained from
quantum-chemical calculations are in very good agreement with
the EXAFS determinations. The calculations also support the higher
APTES affinity for Pb²⁺ compared to Cd²⁺. A novel method to facil-
itate the separation of small organofunctionalized nanoparticles
from the aqueous medium, consisting in preparing multiphase

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Corrigendum

Corrigendum to “EXAFS and DFT study of the cadmium and lead adsorption on modified silica nanoparticles” [Spectrochim. Acta Part A: Mol. Biomol. Spectrosc. 151 (2015) 156–163]

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The authors regret to inform that in Fig. 7 of this article some information was missing. The corrected Fig. 7 is shown below. Authors would like to apologize for the inconvenience caused.

Fig. 7. Optimized structures of APTES-Pb2+, (a) open-conformer and (b) cyclic-conformer, at the CPCM-B3LYP/CEP-121G level of theory. Interatomic distances in Å. Experimental distances obtained by EXAFS are shown in parentheses.