Atmospheric aerosol growth rates at different background station types

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Received: 4 May 2020 / Accepted: 26 October 2020 / Published online: 12 November 2020
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Abstract
Highly time-resolved particle number size distributions (PNSDs) were evaluated during 5 years (2013–2017) at four background stations in the Czech Republic located in different types of environments—urban background (Ústí nad Labem), industrial background (Lom), agricultural background (National Atmospheric Observatory Košetice), and suburban background (Prague-Suchdol). The PNSD data was used for new particle formation event determination as well as growth rate (GR) and condensation sink (CS) calculations. The differences or similarities of these parameters were evaluated from perspectives of the different pollution load, meteorological condition, and regional or long-range transport. The median growth rate ($4 \text{ nm h}^{-1}$) is very similar at all stations, and the most frequent length of growth lasted between 2 and 4 h. Condensation sink reflects the pollution load at the individual station and their connection to the environment type. The highest median, $CS = 1.34 \times 10^{-2} \text{ s}^{-1}$, was recorded at the urban station (Ústí nad Labem), and the lowest ($CS = 0.85 \times 10^{-2} \text{ s}^{-1}$) was recorded at the agricultural station (National Atmospheric Observatory Košetice). Conditional probability function polar plots illustrate the influence of source location to GR. These primary potential emission sources involve traffic, operation of a power plant, and domestic heating.

Keywords Growth rate · Condensation sink · New particle formation · Source location estimation · Pollution load · Background station

Introduction
Atmospheric aerosols are ubiquitous particles, and their presence in the atmosphere contributes to climate change patterns (Kulmala et al. 2004a). Aerosols affect the climate through direct and indirect effects. Atmospheric aerosols can directly scatter and/or absorb solar radiation and thus directly affect the Earth’s radiation balance. The radiation budget is also affected indirectly by aerosols altering clouds’ optical properties and their lifetime, with aerosol acting as cloud condensation nuclei (CCN) (Kulmala et al. 2004b; Pöschl 2005; Yli-Juuti et al. 2011; Stocker et al. 2013). Nonetheless, the role of aerosols in the climatic system still includes uncertainties strongly influencing model simulations (Zhao et al. 2018). One of the uncertainties is caused by the secondary aerosol formation and its consequent growth, called new particle formation (NPF) event. The process of NPF is favored by the presence of sulfuric acid, and low-volatile oxidized organic vapors; on the contrary, the NPF suppressing factor can be a high amount of pre-existing particles in the atmosphere (Dada et al. 2017; Ling et al. 2019). The favoring or suppressing NPF conditions are very closely connected with the behavior of the particle growth rate (GR) and condensation sink (CS).

The GR reflects the sum of all gas-to-particle conversion processes. It contains essential information on the chemical processes that affect growth (Kuang et al. 2012).

The CS values are influenced by the atmospheric pollution load of existing particles and gas precursors. Large CS is a result of the scavenging of freshly nucleated particles and
condensable vapors by existing particles (Zhang et al. 2016). It is still not clear if the high amount of pre-existing particles has a positive or negative effect on NPF, however (Hamed et al. 2007; Nie et al. 2014; Zhang et al. 2016; Zhao et al. 2018; Chu et al. 2019). Since the level of condensation sink is generally higher in more polluted areas as compared to a clean environment (Kulmala et al. 2004a; Pushpawela et al. 2018; Chu et al. 2019), the CS could be marked as an indicator for NPF occurrence (Pikridas et al. 2015; Zhang et al. 2016). Polluted locations, however, contain an abundance of condensable materials resulting in the possibility of observing increased CS along with an elevated number of NPF in these environments (Kulmala et al. 2017; Bousiotis et al. 2019).

The dependency of GR on particle size was recorded (Pushpawela et al. 2018) as well as the relation to pollution loads (Jeong et al. 2010; Bousiotis et al. 2019). Nevertheless, anthropogenic pollution is not the only influencer for GR values. Apart from pollution in cities, other condensation materials supporting GR like volatile organic compounds are produced by nature (Zhao et al. 2018). Synoptic studies (Kerminen et al. 2018; Nieminen et al. 2018) show the influence of natural condensation products on the GR level.

In this work, we have used data from high time-resolved particle number size distribution measurements to investigate any differences or similarities to GR and CS at four background stations located in different types of environments (urban, industrial, agricultural, and suburban) in the Czech Republic, Central Europe. The relations to the meteorological conditions and pollutants’ concentration were compared to find out which mechanisms are crucial at a particular kind of station.

Methods

Measurement sites

The measurements were carried at four background stations in the Czech Republic: Ústí nad Labem-město, Lom, National Atmospheric Observatory Košetice, and Prague-Suchdol (Fig. 1).

Ústí nad Labem-město and Lom stations are located in Ústí nad Labem region belonging to the zone called the “Black Triangle.” The Black Triangle covered the North-Bohemian Brown Coal Basin and parts of Silesia and Saxony. These localities were characterized by low air quality caused by large industrial sources of pollution. After the implementation of environmental laws at the beginning of the 1990s, the air quality started to improve. However, extensive opencast brown coal mining and the presence of large combustion sources and the petrochemical industry still adversely affect the air quality in the region as well as at both measurement sites (Hykyšová and Brejcha 2009).

Ústí nad Labem-město (Ústí n/L, 50° 39’ 39” N, 14° 2’ 35” E, 147 m a.s.l.) is an urban background station situated in a residential area 500 m from the town center. One hundred sixty meters north-east of the station is a road with a traffic volume of 21,000 cars per day (RSD 2020). A chemical company that specialized in synthetic resins, dyes, and chemical products (Ústí 2020) is located 1.3 km west of the station. The population of Ústí nad Labem is 93,000 inhabitants (CSO 2020).

Lom (50° 35’ 8” N, 13° 40’ 24” E, 265 m a.s.l.) is a rural background station. The opencast mine Bílina (1.8 km south-east from the measurement site) and a petrochemical complex (4 km south-west from the station) are located in the vicinity of the station. The complex is focused on a refinery and petrochemical product production (UNIPETROL 2020) and is the largest chemical production facility in the Czech Republic. The population of the nearest small town Lom (approximately 500 m north-west of the station) is 3,700 inhabitants, and the population of Litvinov at distance of 5 km north-west is 24,000 inhabitants (CSO 2020).

National Atmospheric Observatory Košetice (NAOK, 49° 34’ 24” N, 15° 4’ 49” E, 534 m a.s.l.) is a rural background station located in Bohemian-Moravian Highland. The agricultural landscape is composed of forest, fields, and meadows. NAOK is surrounded by several villages within a 3-km distance. Some of them have less than 30 inhabitants; in the largest village Košetice, there are 709 inhabitants (CSO 2020). A major highway D1 with a traffic volume of 40,000 cars per day (RSD 2020) is approximately 7 km NE of NAOK. A factory specialized in wood processing is located 7.5 km west of the sampling site.

Station Prague-Suchdol (Suchdol, 50° 7’ 35.507” N 14° 23’ 4.700” E, 277 m a.s.l.) is a suburban background station located in the campus of the Czech Academy of Sciences in a north-west residential part of Prague. Prague is the capital of the Czech Republic with 1.2 million inhabitants. The population of the residential quarter Suchdol is 5,800 inhabitants (CSO 2020). The traffic volume at the nearest road (200 m north of the site) is around 15,000 cars per day (Kubelová et al. 2015).

Emission inventories

Emission inventories for individual regions of the Czech Republic provided by the CHMI were used for local sources’ identification. This study deals with emissions of PM10 (Ústí n/L and Lom) or total suspended particles—TSP (NAOK and Suchdol), SO2, and NO2.

Since Ústí n/L and Lom are located in the same region, the emission inventory is the same for both stations. The main share of emissions was produced from coal mining and handling, public electricity and heat production, local heating, chemical industry, and traffic. NAOK emissions were dominantly composed of local heating, and traffic and agriculture
activities. Emissions at Suchdol were characterized only by REZZO (Register of Emissions and Sources of Air Pollution) categories, without any possibility to distinguish among the individual sources. Despite this, the emissions of the agglomeration of Prague were mainly produced by local heating, traffic, and industry (CHMI 2019). Detailed graphs are in Supplementary Materials (Fig. S1).

Instrumentation

Particle number size distributions (PNSD) were measured by scanning mobility particle sizers (SMPSs). The details of the instrumentation setting are listed in Table 1. The PNSD data were collected and proceeded according to the standards developed within the Aerosol, Clouds and Trace gases Research InfraStructure (ACTRIS) project. The measurements fulfilled all standard requirements and recommendations (Wiedensohler et al. 2012; WMO 2016). The instruments were regularly calibrated in the European Center for Aerosol Calibration (ECAC): the instruments at Ústí n/L and Lom twice per year, at NAOK, and Suchdol once every 2 years.

Data availability

At three stations (Ústí n/L, NAOK, and Suchdol), PNSD data were available from 2013 to 2017. At the Lom station, data were available only for 2017. The amount of missing data (Table 2) is connected to regular maintenances, calibrations in ECAC, and minor instrumentation failures. The evaluated period for all stations was set to the part of the year when the particle formation probability is higher, i.e., from March to October. Data from late autumn and winter were excluded because of the assumption that photochemical reactions would be suppressed as a result of low values of global radiation.

Auxiliary data

Data from the National Air Quality Network (NAQM) operated by the Czech Hydrometeorological Institute (CHMI) were used to investigate air quality conditions at the stations. At each station, air quality data (sulfuric dioxide (SO2), nitrogen dioxide (NO2), and particulate matter (PM10)) were recorded by automatic analyzers. Also, the basic meteorological parameters (temperature (T), relative humidity (RH), wind speed (WS), wind direction (WD), and global radiation (GLB)) were measured directly at all the stations. The manufacturers and methods of the measurements mentioned above are shown in Table S1 in the Supplementary Materials (SM).

Growth rate, condensation sink, and H2SO4 proxy calculation

Particle growth rate (GR) of freshly formed particles during NPF event days was calculated from the time evolution of the

Table 1 Instrumental setup of PNSD measurement at the individual stations

| Station  | SMPS type          | Condensation particle counter (CPC) type | Size range (nm) | Measurement interval (minute) |
|----------|--------------------|----------------------------------------|----------------|-----------------------------|
| Ústí n/L | Custom-made*       | TSI, 3772                              | 10–800         | 5                           |
| Lom      | Custom-made*       | TSI, 3772                              | 10–800         | 5                           |
| NAOK     | Custom-made*       | TSI, 3772                              | 10–800         | 5                           |
| Suchdol  | TSI 3034**         | TSI, 3010                              | 10–500         | 5                           |

* Made by the Leibnitz Institute for Tropospheric Research
** Reupgraded by the Leibnitz Institute for Tropospheric Research
geometric mean diameter (GMD) values of particles smaller than 100 nm (Jeong et al. 2010):

\[
GR = \frac{\Delta \text{GMD}}{\Delta t}
\]

(1)

where \(\Delta t\) is a time interval from the start to the end of growth and GMD was determined as (Hinds 1999):

\[
\text{GMD} = \exp\sum\frac{(\ln D_{pi}) \times N_i}{\sum N_i}
\]

(2)

where \(D_{pi}\) is the particle diameter of the size bin \(i\), and \(N_i\) is the particle number concentration in the \(i\)th size bin.

Condensation sink (CS) determining the rate of molecule condensation (the loss of molecules) onto the pre-existing aerosol was computed by integrating the PNSD (Kulmala et al. 2004a):

\[
\text{CS} = 2\pi D\beta_M D_p \left( D_p \right) n \left( D_p \right) dD_p
\]

\[= 2\pi D\sum \beta_M D_p N_i
\]

(3)

where \(D\) is the diffusion coefficient, \(D_p\) is the particle diameter, \(\beta_M\) is the transitional regime correction factor, \(n\) is the number concentration, \(D_{pi}\) is the particle diameter of size bin \(i\), and \(N_i\) is the particle number concentration in the \(i\)th size bin.

For the calculation of diffusion coefficient \(D\), we supposed sulfuric acid to be the primary condensing vapor in this region, as two stations (Ústí n/L and Lom) are located in the industrial region. Therefore, \(D\) was computed for sulfuric acid \(D_{H_2SO_4}\) as in (Pushpawela et al. 2018):

\[
D_{H_2SO_4} = \left(5.0032 \times 10^{-6}\right) + \left(1.04 \times 10^{-8}T\right)
\]

\[+ \left(1.64 \times 10^{-11}T^2\right) - \left(1.566 \times 10^{-14}T^3\right)
\]

(4)

where \(T\) is the average temperature in Kelvin at the individual station from March to October during the studied period.

The transitional regime correction factor \(\beta_M\) was calculated according to the Fuchs-Sutugin equation (Fuchs and Sutugin 1971):

\[
\beta_M = \frac{Kn + 1}{1.33\alpha^{-1}Kn^2 + 1.33\alpha^{-1}Kn + 0.38\alpha^{-1}Kn + 1}
\]

(5)

where \(Kn\) is Knudsen number equal to the ratio of the mean free path (\(\lambda = 66\) nm) and the particle diameter \(D_p\) (Hinds 1999); the mass accommodation coefficient \(\alpha\) was assumed to be 1, similarly to for example (Skrabalova et al. 2015).

Sulfuric acid \((H_2SO_4)\) present in the atmosphere in gas phase is the key factor for NPF. H\(_2\)SO\(_4\) in the gas phase is a product of SO\(_2\) oxidation by an OH radical (Zhang et al. 2011). In view of the fact that direct measurement of H\(_2\)SO\(_4\) is not common in Czech Republic, the calculation of proxy\(H_2SO_4\) concentrations by the equation (Petäjä et al. 2009) was used:

\[
\text{proxyH}_2\text{SO}_4 = k_3 \frac{[SO_2] \times GLB}{CS}
\]

(6)

where \(k_3\) is the scaling factor (calculated according to Petäjä et al. 2009), \([SO_2]\) is the \(SO_2\) concentration, GLB is the global radiation, and CS is the condensation sink (calculated from Eq. 3).

**NPF event classification**

Data were analyzed for NPF event occurrence according to the method by Dal Maso et al. (2005). Every day with available data was classified either as new particle formation event day (NPF), non-event day (NON), or undefined event day (UND) (Table 3). The day was classified as NPF if aerosol particle formation began in the nucleation mode (i.e., below 20 nm) and this new mode was recorded for more than 1 h with signs of growing. Only data recorded during NPF event days were used for further analysis.

**Source location estimation**

The potential location of sources influencing the growth rate values was computed in two ways, estimating both nearby and distant sources. Long-range transport was investigated through backward trajectories calculated by HYbrid Single-Particle Lagrangian Integrated Trajectory HYSPLIT\(_4\) model (Draxler and Rolph 2013). Four-day backward trajectories were estimated every 6 h (00, 06, 12, 18 UTC), with Global Data Assimilation System (GDAS) meteorological data with 1° \(\times\) 1° grid resolution used. The height of receptors was set to 500 m AGL. With the trajectories, the potential source contribution function (PSCF) was calculated by TrajStat plugin in MeteoInfoMap software (Wang et al. 2009). The PSCF estimates the probability of the source geographical location in a

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**Table 2** PNDS data availability at the stations

| Station   | Evaluated period       | Missing data (%) | NPF event days (%) |
|-----------|------------------------|-------------------|--------------------|
| Ústí n/L  | March–October 2013–2017| 16                | 33                 |
| Lom       | March–October 2017     | 13                | 40                 |
| NAOK      | March–October 2013–2017| 15                | 33                 |
| Suchdol   | March–October 2013–2017| 30                | 35                 |
grid, with values close to 1 meaning the highest probability. The probability was calculated in $1^\circ \times 1^\circ$ grid, the threshold value was set to median and 75th percentile of the GR, and a weighing function was applied (Zíková et al. 2016). The weighting factor determination is listed in Table S2 in the SM.

The local transport was estimated from the relation between wind speed, wind direction, and a selected variable, obtained by conditional probability function (CPF). The CPF polar plots, illustrating the probability of the occurrence of a concentration at given wind direction and wind speed, were calculated by the R package Openair (Carslaw and Ropkins 2012).

Air mass origin was analyzed from the backward trajectories clustered in the HYSPLIT model, based on the total spatial variance analysis.

## Results

### Basic overview

The atmospheric pollutant concentrations, condensation sink, and growth rate at each station were compared (Table 4). There were visible differences in pollution load between the stations related to the station’s location within the Czech Republic. At the Ústí n/L and Suchdol stations, close to towns, high concentrations of NO$_2$ and PM$_{10}$ and high values of CS were recorded. The highest levels of SO$_2$ and GR were measured at the Lom station. The vicinity of the refinery complex may affect the mentioned values because of emission from the oil products’ processing. The rural background station NAOK reported the lowest amounts of all studied pollutants. Overview of the dependency of GR on selected meteorological elements, air pollutants, proxy H$_2$SO$_4$ concentration, and CS values is plotted in Fig. S2 and Fig. S3 in the Supplemental Materials. Constant GR values are visible over the whole measured range of most variables.

### New particle formation and growth rate

The annual cycle of NPF events did not follow the same pattern at all stations. NPF events frequency peaked in different months at individual stations—in July at Ústí n/L, in May at Lom, in June and August at NAOK, and in August at Suchdol. By contrast, the lowest NPF frequency was observed in October at almost all stations, except at Lom (Fig. 2a). As the Lom results are based on one season, however, it is possible that the same pattern as in other stations would be observed if data over a longer period were available.

The overall median GR was $3.9 \pm 0.3$ nm h$^{-1}$ at all stations in the period under review. The minimal and maximal values of GR ranged from 0.77 to 17.34 nm h$^{-1}$ at Ústí n/L, from 1.97 to 12.49 nm h$^{-1}$ at Lom, from 0.77 to 15.70 nm h$^{-1}$ at NAOK, and from 0.73 to 15.04 nm h$^{-1}$ at Suchdol. Growth rates during NPF events fluctuated in individual months and stations as well. Whereas at Ústí n/L and NAOK the highest median GR was recorded in October, at the Suchdol station, the GR peaked in May and at the Lom station in March. On the contrary, the lowest median GRs were recorded at Lom and NAOK in April, in July at Suchdol, and in May at Ústí n/L (Fig. 2b). These results show relatively weak annual variability with no visible response to various seasons’ conditions.

### Condensation sink daily cycles

Each station was characterized by the daily cycle of CS, connected with pollutants’ loads. At the urban and suburban stations Ústí n/L and Suchdol stations, the connection with morning and evening traffic rush hours was quite pronounced (Fig. 3). In these two cases, increasing CS values after 3 AM peaking at 7 AM (CS~1.7·10$^{-2}$ s$^{-1}$ and 1.2·10$^{-2}$ s$^{-1}$) were observed. After reaching its maximum in the morning hours, CS decreased until the evening rush hour. The secondary maximum observed at all stations between 9 and 12 PM is a result of the decreasing boundary layer level.

At Lom, a different CS daily pattern was found. The decrease of values after morning peak continued only slowly, and a secondary maximum at noon was recorded (Fig. 3). The response of CS to the atmospheric boundary layer (ABL) evolution was most visible at NAOK. CS dropped (9·10$^{-3}$ s$^{-1}$) after sunrise and kept its low values (7.5·10$^{-3}$ s$^{-1}$) until the evening hours (Fig. 3).

The evidence of pollutant influence on CS values is visible through normalized daily cycles of selected pollutants, namely

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**Table 3** The number of classified events in individual months in the evaluated period

| Station | Event/Month | III. | IV. | V. | VI. | VII. | VIII. | IX. | X. |
|---------|-------------|-----|-----|---|----|-----|------|----|----|
| Ústí n/L | NON | 40 | 34 | 26 | 22 | 24 | 18 | 32 | 38 |
| | NPF | 31 | 42 | 46 | 53 | 60 | 55 | 38 | 19 |
| | UND | 70 | 57 | 57 | 45 | 40 | 42 | 65 | 72 |
| | Total | 141 | 133 | 129 | 120 | 124 | 115 | 135 | 129 |
| Lom | NON | 10 | 12 | 7 | 13 | 7 | 5 | 9 | 12 |
| | NPF | 9 | 11 | 16 | 11 | 13 | 10 | 5 | 10 |
| | UND | 12 | 5 | 7 | 6 | 6 | 4 | 8 | 6 |
| | Total | 31 | 28 | 30 | 26 | 19 | 22 | 28 |
| NAOK | NON | 61 | 34 | 25 | 28 | 40 | 32 | 51 | 60 |
| | NPF | 48 | 49 | 48 | 52 | 51 | 49 | 30 | 19 |
| | UND | 43 | 47 | 59 | 58 | 51 | 47 | 40 | 32 |
| | Total | 152 | 130 | 132 | 138 | 142 | 128 | 121 | 111 |
| Suchdol | NON | 19 | 19 | 8 | 16 | 6 | 9 | 12 | 24 |
| | NPF | 36 | 40 | 49 | 47 | 30 | 50 | 34 | 17 |
| | UND | 69 | 49 | 64 | 43 | 39 | 42 | 37 | 65 |
| | Total | 124 | 110 | 122 | 111 | 81 | 100 | 90 | 122 |
The Pearson correlation coefficient between CS and PM$_{10}$ was 0.56, 0.40, 0.50, and 0.67 at Ústí n/L, Lom, NAOK, and Suchdol, respectively. The relation between the PM$_{10}$ and NO$_2$ is expressed by the results of 0.77, 0.84, 0.80, and 0.83 for the same station order.

The diurnal variability of NO$_2$ and PM$_{10}$ and its connection to CS were very similar at Ústí n/L and Suchdol. Two peaks of NO$_2$ and PM$_{10}$ (amplitude 10–30 μg·m$^{-3}$) in the morning and evening hours were observed at these two stations, slightly more pronounced at Ústí n/L. On the contrary, at the Lom station, the response of CS to pollutant concentrations was different when compared to the other stations. Stable CS values did not follow well-developed morning peaks of pollutants (PM$_{10}$=50 μg·m$^{-3}$, NO$_2$=17 μg·m$^{-3}$, SO$_2$=13 μg·m$^{-3}$). The secondary maximum of NO$_2$ and PM$_{10}$ in the evening was similar to the CS as at the rest of the stations. At NAOK, the highest pollutants’ concentrations were measured at night (apart from SO$_2$), and the minimum around noon (PM$_{10}$=15 μg·m$^{-3}$, NO$_2$=5 μg·m$^{-3}$). SO$_2$ daily variability behaved differently to NO$_2$ and PM$_{10}$ at all stations; a prevailing morning peak was recorded at Ústí n/L, Lom, and Suchdol, and morning and evening maxima occurred at NAOK. The SO$_2$ daily cycle fluctuated at evening hours at Suchdol (Fig. 4).

The evolution of the H$_2$SO$_4$ concentrations began by a rapid increase from 3 AM reaching its maximum in the morning hours (Ústí n/L 2.71 × 10$^6$, Lom 4.96 × 10$^6$, NAOK 0.99 × 10$^6$, and Suchdol 1.66 × 10$^6$). Then H$_2$SO$_4$ concentrations started to decline steeply at Ústí n/L and Suchdol, gradually at Lom and NAOK until the evening hours. The boxplots showing the dependence of CS on selected meteorological elements, air pollutants, proxy H$_2$SO$_4$ concentration, and GR values are plotted in Fig. S4 and Fig. S5. CS was found to be constant over the whole range for the majority of the variables. Some increase in CS with increasing concentration is well pronounced for NO$_2$ and PM$_{10}$ at the Ústí n/L and Suchdol stations; such dependence was confirmed also by the Spearman correlation being over 0.5 (Fig. S6 in the SM).

Similarly in the daily cycles at Ústí n/L and Suchdol are probably caused by analogous sources of pollutants such as traffic and heating. The petrochemical factory and opencast possibly affect the pollutant concentrations at the Lom station. The weak daily evolution of the studied variables at NAOK is

| GR (nm·h$^{-1}$) | SO$_2$ concentration (μg·m$^{-3}$) | NO$_2$ concentration (μg·m$^{-3}$) | PM$_{10}$ concentration (μg·m$^{-3}$) | CS$_{H2SO4}$ × 10$^{-3}$ (s$^{-1}$) | Proxy H$_2$SO$_4$ × 10$^6$ (molecules·cm$^{-3}$) |
|-----------------|----------------------------------|----------------------------------|----------------------------------|----------------------------------|----------------------------------|
| **Ústí n/L**    | Median                           | 3.85                             | 4.15                             | 20.40                            | 20.79                            | 13.44                            | 0.65                             |
|                 | Mean                             | 4.39                             | 4.65                             | 21.55                            | 22.77                            | 14.65                            | 1.58                             |
|                 | SD                               | 2.23                             | 2.81                             | 7.45                             | 9.72                             | 8.11                             | 2.51                             |
| **Lom**         | Median                           | 4.20                             | 5.77                             | 7.83                             | 22.04                            | 9.40                             | 2.36                             |
|                 | Mean                             | 5.08                             | 6.62                             | 7.94                             | 23.68                            | 10.31                            | 5.11                             |
|                 | SD                               | 2.54                             | 4.33                             | 3.52                             | 11.58                            | 6.27                             | 2.41                             |
| **NAOK**        | Median                           | 3.62                             | 1.28                             | 6.14                             | 14.29                            | 8.54                             | 0.63                             |
|                 | Mean                             | 4.13                             | 1.60                             | 6.55                             | 15.77                            | 9.14                             | 0.87                             |
|                 | SD                               | 2.24                             | 1.15                             | 2.53                             | 6.97                             | 4.51                             | 0.83                             |
| **Suchdol**     | Median                           | 3.98                             | 2.79                             | 14.89                            | 18.33                            | 11.52                            | 1.01                             |
|                 | Mean                             | 4.47                             | 3.36                             | 17.82                            | 20.46                            | 12.99                            | 1.48                             |
|                 | SD                               | 2.30                             | 2.00                             | 9.66                             | 9.78                             | 7.67                             | 1.62                             |

Table 4  Median, mean, and standard deviation (SD) of GR, air pollutants’ concentrations, values of CS, and proxy H$_2$SO$_4$ at Ústí n/L (N = 341), Lom (N = 86), NAOK (N = 346), and Suchdol (N = 302) in the period March–October 2013–2017.
likely a result of the station’s location. Its position far from significant pollution sources enables the observation of the response of pollutants’ concentrations to ABL height development during the day, especially the dilution effect to atmospheric components.

**Growth rate patterns**

Three categorized characteristics were used to describe the properties of GR at each station—the time of the beginning of the growth, length of the growth, and growth rate. The data were split into intervals (bins) representing a 2-h (or 2 nm h$^{-1}$) period starting at zero. A particular bin contains the number of recorded occurrences of a specific range. The number of a score in the individual interval was divided by the total number of incidences for better comparison between the stations.

Some common patterns were found across the stations—the most frequent start time of the growth was from 10:00 to 12:00 UTC. However, individual behavior was observed as well. An increase of the frequency of the particle growth start time was recorded from 16:00 to 18:00 at Suchdol, and from 18:00 to 20:00 at Lom. At NAOK and Ústí n/L, the frequency of the GR start time continuously declined from its maximum between 10:00 and 12:00 (Fig. 5a).

Similarly to the time of the beginning, also the length of growth showed both similarities and differences at individual stations. The most typical length of growth was 2–4 h at all stations. The second highest frequency was recorded in length category 0–2 h, except for NAOK station with the 4–6 h being more frequent. Generally, at NAOK, longer lengths of growth were observed more often—growth frequently lasted up to 10 h, and NAOK was also the only station where growth took more than 12 h with a frequency of over 5% (Fig. 5b). The low value of CS may influence the GR observations. In a clean environment, such as at NAOK, existing particles do not inhibit the already-initiated growing process. In agreement with the median GR value close to 4 nm h$^{-1}$ at all stations, the highest frequency of GR was in categories 2–4 and 4–6 nm h$^{-1}$. At almost all stations, a gradual decrease in GR frequency in the direction toward upper classes was recorded. The GR frequency at Lom was different from other stations with some increase in GR over 10 nm h$^{-1}$ (Fig. 5c).

**Local, regional, and long-ranged GR sources’ identification**

**Wind speed and direction dependence**

CPF polar plots were used to represent the location of potential sources, mostly influencing the GR. Local and regional sources were estimated for events with a GR higher than the 75th percentile at the individual station, and the same analysis was also done for SO$_2$, NO$_2$, PM$_{10}$, proxy H$_2$SO$_4$ concentrations, and CS values. For better orientation in the results, dashed red lines were added into polar plots, showing the spatial sector, where the probability of GR reaching 75th percentile is over 0.5.

At Ústí n/L, the strongest signal for the GR was identified from NNE wind directions with a wind speed of 4–6 m s$^{-1}$. The probability of other variables in this direction is under 0.4. In contrast, a SO$_2$ probability close to 1 was observed in the NE wind direction. In the NE direction, there is a crossroad with high traffic density and in the SE (GR second strongest signal) there is a power plant.
The strongest CPF signal at Lom was found from NE to SE directions. An opencast Bílina, a power plant, and two companies focused on glass production are at these locations. The results of the other studied pollutants do not correspond with the GR polar plot results. The probability of GR around 0.4 in the SW direction may be connected with the proximity of chemical factories which is confirmed by the SO$_2$ and proxy H$_2$SO$_4$ results.

At NAOK, a probability close to 0.9 was recorded during NE winds with 8–10 m s$^{-1}$ wind speed that is indicative of a regional source. One of the main Czech highways, D1, is located in the NE direction. The rest of the studied variables show probability over 0.3 associated with the NE direction, especially the PM$_{10}$ and proxy H$_2$SO$_4$ results, which were between 0.6 and 0.8. However, a stronger signal was observed for SO$_2$, CS, and proxy H$_2$SO$_4$ in the E direction.

At the Suchdol station, a GR probability over 0.5 was recorded in some locations from NW to NE. These locations are sources like a road with high traffic density and local heating. The results of the other analyzed components not completely agree with the GR outcomes, but the SO$_2$, PM$_{10}$, CS, and proxy H$_2$SO$_4$ signals were over 0.4 in NE wind direction. The highest concentration of NO$_2$ was associated with eastern winds with a low wind speed, which can be related to local sources. A proxy H$_2$SO$_4$ probability above 0.6 was recorded in the SW and S directions, which differs from the rest of the results (Fig. 6). Despite some similarities, it was not found that the GR at all stations is probably entirely affected by the same sources as the studied variables. Thus, either some additional pollutants would have to be considered or some of their combinations. Although the stations Ústí n/L, Lom, and Suchdol are close to emission sources, our results indicate that direct emissions of local sources do not affect GR substantially. Presumably, fresh emissions need to undergo chemical reactions to be effective for increasing GR. This assumption may be supported by the results from NAOK, where proxy H$_2$SO$_4$ showed a higher probability of occurrence (0.6) in the same direction as GR. NAOK is distant from direct pollution sources, unlike the other three stations.

Potential source location

The potential source contribution function (PSCF) analyses were done for a critical value set to the individual 75th percentile of GR at each station. Generally, similar regional sources and influence of long-range transport across the stations were found.

Western sources located in line from Central France to Germany influence the GR values at Ústí n/L, Lom, and NAOK (probability 0.3–0.6). Only at Suchdol was no influence by western long-range transport seen (Fig. 7). The potential location of sources for all stations was identified in the north part of Italy, and also from the SE (probability 0.3–0.6). Suchdol and Lom were affected by regional/long-range transport from the NE through Poland.

The different probability of source contribution at Lom as compared to the rest of the stations was observed by eastern sources located in Poland, Austria, Hungary, and the Slovak Republic. Higher probabilities in the Lom results can be caused by a low number of trajectories, short data period, and also a small number of GR values, however.

Within the space of the Czech Republic, GR was influenced by SW, SE, and E sources. The east part of the Czech Republic is strongly influenced by heavy industry. Industrial complexes specialized in coal and steel processing are located close to the Czech-Polish border. Other industrial regions in the Czech Republic did not influence the GR values at the studied stations.

Growth rates in different meteorological conditions and pollution loads

The relation of the rate of growth and the atmospheric conditions was investigated based on the split dataset of all variables into two parts—events with GR below the 25th and above the 75th percentile (Fig. 8). The differences in the meteorological parameters or pollutants’ concentrations in the two datasets were tested by the Mann-Whitney $U$ non-parametric test. No statistically significant difference was found for the meteorological variables.
(global radiation, wind speed and wind direction, and temperature) except global radiation at Ústí n/L. These results indicate that the observed meteorological conditions probably do not strongly influence the GR values. The meteorological conditions were similar when GR was below the 25th and above the 75th percentiles. Two cases representing differences between the two groups in pollutants’ concentrations were recorded for SO₂ at Lom and

Fig. 6 CPF polar plots of the probability of concentrations exceeding the 75th percentile of GR and hourly SO₂, NO₂, PM₁₀ concentrations, CS, and proxy H₂SO₄ values connected with the wind direction and wind speed at Ústí n/L, Lom, NAO Košetice, and Suchdol. Dashed red lines indicate the spatial sector in the polar plots where the GR results exceeded 0.5 probability.

Fig. 7 PSCF of the sources calculated for threshold set to 75th percentile of GR values. The black point indicates the station’s position—Ústí n/L, Lom, NAO Košetice, and Suchdol.
for NO2 at Suchdol station. These results indicate that the observed meteorological conditions do not strongly influence the GR values. Although the pollutants’ concentrations were usually higher in the GR > 75th percentile dataset (apart from SO2 at NAOK and Suchdol and proxy H2SO4 at Ústí n/L, NAOK, and Suchdol), there was probably an additional (not observed) component (e.g., VOC, OH radical, condensable vapor concentration, aerosol chemical composition (O’Dowd et al. 2002; Petäjä et al. 2009; Zhang et al. 2011; Kulmala et al. 2017)) in the atmosphere that controls the growth rate.

The results from the Spearman correlation coefficient applied on the two datasets did not show any similarities across the stations. The strength between variables differs from station to station, and no general conclusion is evident (Fig. S6 in SM).

Fig. 8 A statistical overview of meteorological parameters and pollutants’ concentrations during events when the measured GR was below the 25th (columns without outline borders), or above the 75th percentile values (columns with black outline borders), the top of the column is the median. The error bars show the 25th and 75th percentiles. GLB, global radiation; T, temperature; RH, relative humidity; WS, wind speed; SO2, sulfur dioxide; NO2, nitrogen dioxide; PM10, particulate matter; CS, condensation sink; H2SO4, H2SO4 proxy

Fig. 9 Statistical cluster analysis of air mass backward trajectories showing the dependency of GR on different air mass origin at Ústí n/L, NAOK, and Suchdol. One receptor site, Suchdol, was chosen, as the distance between stations is smaller than 100 km, the spatial resolution of the model. Cluster number 1—fresh continental; 2—fresh marine; 3—aged continental; 4—fresh marine; and 5—continental air mass. The boxes are colored the same as clusters, the black horizontal line is the median, the borders of the boxes show the 25th and 75th percentiles, and the error bars indicate the minimum and maximum values.
As a proxy for aerosol composition (that was not available for most of the stations), the air mass origin was considered and the GR was compared in various air masses. As the HYSPLIT model resolution was smaller than the geographical distance of the stations, only one station was selected as a receptor site—the Suchdol station, which represents the middle of the Czech Republic—for the clustering. For the air mass analyses, only days with the NPF occurrence at all stations were included in the calculation (53 days). The Lom station was excluded for this part because of the lack of data (only 6 days). The total amount of 198 backward trajectories were clustered to 5 clusters. Cluster numbers 1, 3, and 5 were of continental origin (number 3 is aged air mass), and clusters 2 and 4 are of fresh marine origin (Fig. 9). Throughout all stations, the results of GR in different air masses were similar. Cluster number 1 included only several days, so the statistics are not very good. The highest GR median values were observed in clusters 5 (continental), and 2 (fresh marine), apart from Ústí n/L, where in cluster 3 (aged continental) the median GR was slightly higher. The lowest median GR occurred in cluster 4 (fresh marine). Thus, no clear signal in the connection between the GR and continental or maritime air masses, with different chemical composition, that would be applicable for all three stations was found.

**Summary and discussion**

The goal of the study was to investigate differences and/or similarities influencing the formation of aerosol particles and their consequent growth in different types of the background environments. This evaluation is based on GR and CS characteristic during NPF events.

Although NPF events’ occurrence was considered to be more frequent in clean areas (Kulmala et al. 2017; Kerminen et al. 2018), these findings do not apply to the GR and CS values. The observation in megacities proved frequent NPF events, which is not in line with theory. Another physical and chemical mechanisms influence this process in a mixture of urban pollution (Kulmala et al. 2017). Similar results were found in our data; NPF occurred also at the most polluted station Lom. According to review by Kerminen et al. (2018), less polluted areas are typical for lower GR compared to highly polluted locations; the contrary is true for CS. An abundance of condensable materials, manifested by higher CS values, probably causes increased GR in polluted areas (Bousiotis et al. 2019; Chu et al. 2019).

At our background stations, the differences in the CS daily evolution reflecting the type of environment were found. The response to anthropogenic activities (morning and evening rush hours) was recorded at Ústí n/L and Suchdol; unique CS evolution, probably connected to close mining activities, occurred at Lom. At the cleanest location, i.e., NAOK, the linkage between pollution dilution and ABL daily development was visible.

The calculated GR values referring to a median value around 4 nm h−1 correspond to the results of similar stations type reported in recent studies (Bousiotis et al. 2019; Kerminen et al. 2018; Nieminen et al. 2018). A unique GR length time lasting more than 10 h with frequency over 10% was recorded at NAOK.

Potential local and regional sources influencing GR were associated with directions where pollution sources are situated. The links with the sources of other studied pollutants were confirmed. The proxy H2SO4 at NAOK suggests that very close direct emissions may need to undergo a chemical reaction with other atmospheric components to act on GR. Only at this station, thanks to the absence of direct pollution sources, was it possible to identify the same source site of the studied variables (in particular for PM10, CS, and proxy H2SO4) as for GR. Alternatively, some other atmospheric species probably affect GR. Regional and long-range transport seem to influence GR similarly.

The investigated similarities and differences at the stations reflect the influence of the type of surrounding environment, local topography, and the character of the emissions. Although all stations are background, we can observe the same patterns that are usually more pronounced in the more polluted areas. Our results show that among background stations, the found differences can be related to specific pollutants and these findings can be helpful for other comparisons and further studies of NPF events.

**Supplementary Information** The online version contains supplementary material available at https://doi.org/10.1007/s11356-020-11424-5.

**Acknowledgments** This work was supported by the Ministry of Education, Youth and Sports of the Czech Republic, within the project for support of the national research infrastructure ACTRIS—participation of the Czech Republic (ACTRIS-CZ-LM2018122). Proofreading was done by Mr. Sean Mark Miller, Ph.D.

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**Funding** This work was supported by the Ministry of Education, Youth and Sports of the Czech Republic, within the project for support of the national research infrastructure ACTRIS—participation of the Czech Republic (ACTRIS-CZ-LM2018122).

**Data availability** The datasets generated during and/or analyzed during the current study are available from the corresponding author on reasonable request.
Compliance with ethical standards

This manuscript has not been published and is not under consideration for publication elsewhere. All the authors have read the manuscript and have approved this submission.

Ethical approval Not applicable

Consent to participate Not applicable

Consent to publish Not applicable

Competing interests The authors declare that they have no conflict of interest.

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