Effects of Heavy Metal Contamination from an Abandoned Mine on Tomato Growth and Root-knot Nematode Development

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Physicochemical characteristics and heavy metal content of soils located along the drainage way of an abandoned mine at Busan, Korea (35°31'N, 129°22'E) (contaminated soil; CS) and uncontaminated soils (50-70 m apart from the drainage way) (NS) were examined. Growth of tomato plants (Solanum lycopersicum cv. Rutgers) in CS and NS, development of the root-knot nematode (Meloidogyne incognita) as root-knot gall formation on tomato plants, and non-parasitic nematode populations in soil were also examined. Growth of tomato plants, root-knot gall formation, and non-parasitic nematode populations in soil were significantly reduced in CS with higher As content, lower pH, higher electrical conductivity (EC), and lower available phosphate (av. P2O5) than in NS. None of the other physicochemical characters examined differed significantly between CS and NS (low and no significance) and were above or below the critical levels detrimental to plant growth and nematode development, suggesting that As may be the primary hazardous heavy metal in CS. The toxicity of As might be enhanced at low pH in CS because exchangeable forms of some heavy metals increase with the decrease of soil pH. The heavy metals, especially As, may have contributed to increasing EC and decreasing av. P2O5. Therefore, the effects of mine drainage contamination from the abandoned mine were derived primarily from contamination by heavy metals such as As. These may have been enhanced in toxicity (solubility) by the lowered pH, increased soil salinity (EC) and decreased av. P2O5. Our results suggest synergistic adverse effects on the plant and the nematode by decreasing osmotic potential and nutrient availability.

Keywords: gall formation, heavy metals, Solanum lycopersicum, Meloidogyne incognita, mine drainage

Abandoned mines are sources of heavy metals that contaminate nearby soil and water systems via mine drainage and debris containing a large quantity of heavy metals, which are neither decomposed to non-toxic compounds nor transformed into safe material (Adriano, 1986; McBride, 1994; Rajkumar et al., 2005). Heavy metal contamination affects soil health by inhibiting soil functioning due to decreases in microbial activity, soil enzymes, nitrogen fixation, and growth of microfloral communities such as fungi, algae, and photosynthetic bacteria (Mhatre and Pankhurst, 1997). Accumulation of heavy metals such as Cu, Pb, Zn, and As in soil can lead to plant physiological disorders and even death, although damage from these heavy metals may not be very high (Holmgren et al., 1993; Vulava et al., 1997). For example, germination of pine seeds is reduced in proportion to the heavy metal content (Seo et al., 2006). Moreover, Cd and Hg accumulate in crop plants and are detrimental to human and animal health (Holmgren et al., 1993; Vulava et al., 1997).

Soil faunal communities such as nematodes, protozoa, and earthworms are also reduced in abundance and richness in soils contaminated with heavy metals (Mhatre and Pankhurst, 1997). In particular, all nematode trophic groups, including bacterivores, fungivores, predators, and plant parasites, are reduced in abundance and richness (Park, 2011). In soil ecosystems, a reduction in all nematodes other than plant parasites indicates a reduction in soil functioning and health, as non-parasitic nematodes play an important role in the soil ecosystem, contributing directly to nitrogen mineralization and biomass distribution and indirectly through grazing on decomposer microbes, secreting ammonium, and immobilizing nitrogen in live biomass (Beare, 1997; Ferris et al., 1998; Ingham et al., 1985; Neher, 2001). On the other hand, a reduction in plant parasitic nematodes due to heavy metal contamination implies an improved soil health, defined as the continued capacity of soil for biological productivity, plant health, and productivity (Doran et al., 1996).

Few studies have been conducted on the influence of
heavy metals on soil functioning and plant health, especially in relation to diseases caused by plant-parasitic nematodes. The present study examined the effects of heavy metals on the growth of tomato and the development of the root-knot nematode, *Meloidogyne incognita*, using soils sampled from areas around an abandoned mine in Busan, Korea. The physicochemical characteristics of the soil (including heavy metal content) were analyzed, and tomato seedlings planted in this soil and plants inoculated with *M. incognita* were examined for growth and nematode disease development. Population changes in non-parasitic nematodes in the soil were also examined.

**Materials and Methods**

**Soil sampling and analysis of soil characters.** Soils were obtained from six experimental sites [three sites contaminated with mine drainage (CS: CS-1, CS-2, CS-3) located along the drainage way and three uncontaminated sites with mine drainage (NS: NS-1, NS-2, NS-3) (50–70 m apart from the drainage way) around an abandoned mine at Gijang county, Busan, Korea (35°31’N, 129°22’E) (Fig. 1). This mine was developed for copper and silver mining in 1930 but has been abandoned since 1980. Soil samples of about 1 kg were taken from each site, dried under shade for 2 weeks, and sieved through a 2.0-mm aperture sieve before analysis. Soil texture was determined by measuring the amount of clay, silt, and sand in a 10-g sample suspended in a soil dispersant (using a pipette) after removal of organic matter by boiling in H2O2 for 1 week, and of Ca using HCl (Klute, 1986). Following the methods described by the NIAST (2000), soil pH and electrical conductivity (EC) were analyzed using pH and EC meters, organic matter by the Tuurin method, total nitrogen (T-N) by the Kjeldahl method, available phosphate (av. P2O5) by the Lancaster method using a spectrophotometer, and NO3-N and NH4-N extracted in 2 M KCl using an automatic ion analyzer. Soluble cations (Ca2+, Na+, K+, and Mg2+) were determined by flame photometry, and heavy metal content was determined by ICP atomic emission spectrophotometry using 10% soil solutions containing 0.1 N HCS that were held in a glass flask at 30°C for 1 h and then filtered through Whatman No. 2 filter paper (MOE, 2005).

**Plant growth and nematode development in soils from the abandoned mine.** Tomato (*Lycopersicon esculentum* cv. Rutgers) seedlings (susceptible to the root-knot nematode) were planted and grown in sterilized vermiculite for 3 weeks and then transplanted into 12-cm-diameter clay pots each containing 650 g soil from the experimental sites. Sandy loam culture soil, with a texture similar to the experimental soils, was used as a control and inoculated with or without the root-knot nematode. Each plant was inoculated with 1,000 eggs of *M. incognita* obtained from greenhouse-grown tomato plants used for nematode multiplication. Each treatment was replicated five times. Pots were arranged in a randomized block design on a bench in a greenhouse at 25 ± 2°C and were watered daily to field capacity. About 8 weeks after nematode inoculation, plants were carefully uprooted from pots, and plant shoot height, root length, and weight were measured. Severity of root galling on tomato plants was assessed on a 0–5 rating scale according to the number of galls per plant, as follows: 0 = none, 1 = 1–2, 2 = 3–10, 3 = 11–30, 4 = 31–100, and 5 = > 100 galls (Taylor and Sasser, 1978). The nematode population was extracted from 100-g subsamples of the soil samples and sieved through 30-mesh, 200-mesh, and 325-mesh sieves, followed by the Baermann funnel procedure for 48 h (Southey, 1986). Recovered organisms including nematodes were preserved in triethanolamine-formalin (TAF) fixative at 80°C, mounted on glass slides (Southey, 1986), and observed under a stereoscope microscope to enumerate the nematode populations. All trophic states, excluding second stage juveniles of *M. incognita*, were determined based on stoma and esophageal morphology, and combined and counted as non-parasitic nematodes.

**Statistical analysis.** For statistical analysis of the experi-
mental results, analyses of variance were carried out using SAS version 4.2 (SAS Institute Inc., Cary, NC, USA), with Duncan’s multiple range test (DMRT) to examine significant differences ($P \leq 0.05$) among treatments. Correlation coefficients were computed to test significant correlations between physicochemical soil characters and their influence on plant growth, gall formation by the root-knot nematode, and populations of non-parasitic nematodes, at $P \leq 0.05$ and $P \leq 0.01$.

Results

Soil characteristics. The soil texture of all experimental sites was sandy loam, composed of 53.5–63.6% sand, 2.6–4.6% clay, and 33.8–42.0% silt (data not shown). Other soil characters, except heavy metal content, are shown in Table 1. pH was lower in CS (pH 2.5–2.8) than NS (pH 4.1–4.6); EC was higher in CS (4.80–7.26 dS/m) than NS (0.39–0.46 dS/m); av. P$_2$O$_5$ was lower in CS (0.68–0.77 mg/kg) than NS (4.17–7.98 mg/kg); and nitrogen content [total nitrogen (T-N), NO$_3$-N, and NH$_4$-N] and soluble cations (K$^+$, Ca$^{2+}$, Mg$^{2+}$, Na$^+$) were similar among the CS and NS soil samples, with no significant differences between CS and NS.

The content of As was remarkably higher in CS (especially CS-1 and CS-2: over 5,000 mg/kg) than in NS (Table 2). The Cu content was generally higher in CS than NS except for NS-3, which had the highest level; contents of Cr, Ni, and Zn were somewhat higher in NS than CS; Pb was similar in CS and NS, except in NS-3 which had the highest level.

| Soil characteristics          | CS             | NS             | CS-1 | CS-2 | CS-3 | NS-1 | NS-2 | NS-3 |
|------------------------------|----------------|----------------|------|------|------|------|------|------|
| pH (1:5)                     | 2.5 ± 0.14c    | 2.7 ± 0.07c    | 2.8 ± 0.14c | 4.6 ± 0.28a | 4.1 ± 0.14b | 4.1 ± 0.14b |
| EC (dS/m)$^b$                | 7.26 ± 0.23a   | 7.42 ± 0.18a   | 4.8 ± 0.14b | 0.39 ± 0.01c | 0.46 ± 0.01c | 0.42 ± 0.01c |
| Total N (mg/kg)              | 400 ± 141.4a   | 400.0 ± 141.4a | 300.0 ± 77.8a | 700.0 ± 212.1a | 400.0 ± 0.0a | 500 ± 141.4a |
| NO$_3$-N (mg/kg)             | 0.22 ± 0.01b   | 0.21 ± 0.01b   | 0.2 ± 0.01b | 0.23 ± 0.01b | 0.41 ± 0.16a | 0.31 ± 0.08ab |
| NH$_4$-N (mg/kg)             | 1.25 ± 0.06ab  | 1.26 ± 0.01ab  | 1.25 ± 0.01ab | 1.07 ± 0.01b | 1.35 ± 0.14a | 1.17 ± 0.16ab |
| Av. P$_2$O$_5$ (mg/kg)       | 0.77 ± 0.17d   | 0.69 ± 0.00d   | 0.68 ± 0.01d | 4.17 ± 0.01c | 7.98 ± 0.03a | 6.47 ± 0.04b |
| K [Ex. cation (cmol/kg)]     | 0.12 ± 0.00a   | 0.11 ± 0.03a   | 0.12 ± 0.01a | 0.12 ± 0.00a | 0.11 ± 0.01a | 0.11 ± 0.03a |
| Ca [Ex. cation (cmol/kg)]    | 0.32 ± 0.01abc | 0.46 ± 0.16abc | 0.49 ± 0.15c | 0.43 ± 0.01abc | 0.25 ± 0.01bc | 0.22 ± 0.01c |
| Mg [Ex. cation (cmol/kg)]    | 0.32 ± 0.00bc  | 0.46 ± 0.13ab  | 0.49 ± 0.02a | 0.43 ± 0.03ab | 0.25 ± 0.03c | 0.22 ± 0.01c |
| Na [Ex. cation (cmol/kg)]    | 0.12 ± 0.01a   | 0.11 ± 0.00a   | 0.12 ± 0.00a | 0.12 ± 0.01a | 0.11 ± 0.00a | 0.11 ± 0.00a |

$^a$Figures are averages and standard deviations of two replications. The same letters in a row denote no significant difference at $P \leq 0.05$ by Duncan’s multiple range test (DMRT).

$^b$EC: electrical conductivity

Table 2. Heavy metal contents in soils contaminated with mine drainage (CS: CS-1, CS-2 and CS-3) and nearby non-contaminated soils (NS: NS-1, NS-2 and NS-3)

| Heavy metal | CS-1 (mg/Kg) | CS-2 (mg/Kg) | CS-3 (mg/Kg) | NS1- (mg/Kg) | NS-2 (mg/Kg) | NS-3 (mg/Kg) |
|-------------|--------------|--------------|--------------|--------------|--------------|--------------|
| As          | 5,510.0 ± 193.3a | 5,161.7 ± 247.5b | 2,552.3 ± 13.7c | 36.7 ± 1.9e  | 40.1 ± 1.1e  | 380.0 ± 14.6d |
| Cu          | 67.2 ± 0.5d   | 158.1 ± 1.8b  | 95.9 ± 0.7c   | 25.3 ± 2.3e  | 30.1 ± 5.5e  | 186.9 ± 0.1a  |
| Cr          | 4.2 ± 0.0c    | 2.0 ± 0.0d    | 4.4 ± 0.2c    | 11.5 ± 0.3b  | 11.4 ± 0.1b  | 18.4 ± 0.1a   |
| Ni          | 1.2 ± 0.1c    | 0.2 ± 0.1d    | 1.3 ± 0.1c    | 5.9 ± 0.3b   | 5.6 ± 0.3b   | 8.8 ± 0.1a    |
| Pb          | 20.1 ± 2.0c   | 38.8 ± 0.8b   | 30.4 ± 0.3c   | 27.5 ± 0.8c  | 31.9 ± 2.7c  | 178.8 ± 15.4a |
| Zn          | 45.4 ± 0.3c   | 46.1 ± 0.2c   | 47.0 ± 0.4c   | 62.5 ± 2.7b  | 62.1 ± 0.4b  | 118.3 ± 1.7a  |

$^a$Figures are averages and standard deviations of two replications. The same letters in a column denote no significant difference at $P \leq 0.05$ by Duncan’s multiple range test (DMRT).
control soils with or without nematode inoculation compared to the experimental soils; however, there were no significant differences in plant growth between control soils with nematode inoculation (Control-A) and without nematode inoculation (Control-B) (Table 3).

Gall formation on tomato roots, as a method to examine nematode development, was significantly lower in CS compared to NS, with a significant reduction in nematode populations in CS in the order CS-1, CS-2, and CS-3, which was similar to gall formation.

Correlation of soil characteristics with plant growth, gall formation, and non-parasitic nematode populations. Significant or highly significant correlations were found in plant growth, gall formation, and non-parasitic nematode populations in relation to the soil characters of pH, EC, NO₃⁻-N, av P, P₀₂, and content of the heavy metals As, Cr, Ni, and Zn (Table 4). Most significant decreases in plant growth, gall formation, and non-parasitic nematode populations occurred in relation to As, followed by pH and EC.
Discussion

Growth of tomato plants (shoot height, shoot weight, root length, and root weight) was reduced greatly in CS with lower pH, higher EC, lower av. P₂O₅, and especially with high As compared to NS. The reduction in plant growth was highly correlated with the As content in the soils examined. In this study, the tomato plants grown in CS, especially in CS-1 and CS-2 with As of over 5,000 mg/kg, showed leaf yellowing and wilting 4 weeks after planting, similar to the symptoms of As toxicity found by Machlis (1941), who reported leaf wilting followed by root discoloration and leaf tip necrosis. Reduced growth was also reported for Arabidopsis thaliana cv. Columbia grown in culture media with a high content of Pb (Lee, 2010). In addition, Zn toxicity to crop plants was evident in fields contaminated with over 300 mg/kg of Zn from mining and smelting activities (Chaney, 1993; Marschner, 1995). However, apart from As, other heavy metals such as Pb and Zn were either higher in CS or in NS, with no remarkable differences between the two soils, and were correlated negatively or positively with plant growth, suggesting that these heavy metals had no critical effects on plant growth in our study.

Arsenic (As, atomic number: 33) is regarded as a heavy metal, despite being a semimetal (metalloid), according to the definition of heavy metals associated with contamination and potential toxicity or ecotoxicity (Duffus, 2002). It is a notorious toxin to many organisms, and has been used in pesticides, insecticides (probably toxic to nematodes), fungicides, and herbicides (Abernathy, 1983; Azcue and Nriagu, 1994). In our study, gall formation caused by the root-knot nematode was reduced significantly in tomato plants grown in CS (especially in CS-1 and CS-2 with high As content) compared to that in NS and the inoculated control (Control-A). In addition, the free-living nematode populations in the soil were significantly reduced in CS compared to NS. This result indicates that the reduction in gall formation may be due to direct heavy metal toxicity as an insecticide (nematicide) to second stage juveniles of the root-knot nematode, preventing nematode penetration into root tissues, and to third and fourth stage juveniles or adults inside root tissues, where they induce giant cells (and gall formation) through feeding (Moon et al., 2010). Heavy metals at high concentrations are known to have strong adverse effects on nematode communities in soils (Sánchez-Moreno and Navas, 2007) and accumulate in plant tissues at concentrations hazardous to human health (Kim et al., 1998, 2002). Toxicity to nematodes may also be attributable to the inhibitory effect of As as an herbicide, reducing tomato root growth, consequently inhibiting the formation and development of giant cells related to the gall formation of the root-knot nematode (Son et al., 2009). In our study, the reduction in nematode populations was not as large as the reduction in gall formation, suggesting that the nematicidal activity of As may be lower than its phytotoxic (herbicidal) activity. In this respect, the use of As as a nematicide is not recommended because of its high phytotoxicity.

Correlation coefficients for soil characteristics with plant growth, gall formation, and nematode populations also showed that soil pH, EC, and possibly av. P₂O₅ were nearly the same (P ≤ 0.01) as those with As. The chemical forms of heavy metals in soil and their uptake rates by plants are affected by soil pH, resulting in an increase in exchangeable forms of heavy metals at low pH, as low soil pH results in a concomitant increase in exchangeable forms of Cd, Zn, and Pb (Xian and Shokohifard, 1989). A reduction in av. P₂O₅ in CS can also be explained by its precipitation in heavy metal-contaminated acid soils because av. P₂O₅ precipitates with coexisting Al and Fe hydrolytic metal species, which also decrease soil pH (Ro and Cho, 2000). At the same time, reduced soil pH also results in a concomitant increase in exchangeable cations. In this sense, lower pH in CS than in NS may enhance the toxicity of As and other heavy metals such as Cu whose content was not even lower in CS than in NS. Soil EC, a measure of soil salinity, adversely affects plant growth through decreased osmotic potential (i.e., increased osmotic pressure) and/or the specific effects of individual ions that contribute to soil salinity (Bernstein, 1975). In our study, there were no remarkable differences in the ions examined, except the heavy metal As, that would account for the difference in EC between CS and NS, suggesting that As and other heavy metals that ionized at lower pH (especially Cu) in CS may have increased EC and thus adversely affected plant growth and root-knot nematode development. In conclusion, the effects of soil contamination by drainage from an abandoned mine in our study were derived primarily from contamination by heavy metals (especially As). This contamination led to enhanced toxicity (solubility) by the lowered pH, and to increased soil salinity and decreased av. P₂O₅, thus having synergistic adverse effects on plant growth by decreasing osmotic potential and increasing nutrient deprivation.

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