Tree Structural Diversity and Yield Prediction Models for Tree Species in Old Oyo National Park, Nigeria

Adesoji Akinwumi Adeyemi*, Hussainat Taiwo Taofeek

Department of Forest Resources Management, Faculty of Agriculture, University of Ilorin, Ilorin, Nigeria

Email address:
adeyemi.aa@unilorin.edu.ng (A. A. Adeyemi), adeyemiadesoji@yahoo.com (A. A. Adeyemi)

*Corresponding author

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Abstract: Evaluation of structural and species distribution is essential for conservation work. Besides, reliable baseline information is crucial for a sustainable forest management. We assessed forest structure and formulated yield-prediction models for tree species in Old Oyo National Park (OONP). Systematic sampling technique was adopted for the study. Thirty transects of 1000 m-long were laid at 1 km intervals. Four 0.25ha-plots were then alternately laid at 250 m intervals along each transect, and 120 plots were used for the study. Structural parameters including diameters at the base, breast height, middle, merchantable top and tree heights were measured for trees with Dbh ≥10 cm to compute stand density and volume. Data were analysed using descriptive statistics and regression. Species diversity and richness indices were computed using Shannon-Wiener and Margalef’s Indices. Models were assessed using $R^2$ and standard error of estimate (SEE), overall significance of each model was evaluated using F-test. Model validation was done using t-test and bias. Sixty-four tree species in 23 families were encountered with richness and diversity indices of 1.189 and 3.544, respectively. Only one endangered (Pterocarpus erinaceus) and three vulnerable (Afzelia africana, Vitellaria paradoxa and Maranthes agnesis) species were encountered. Burkea africana was the most locally-abundant in the area with a relative density and basal area of 9.3 and 3.058 m$^2$/ha, respectively. Fabaceae was the most-represented family. The best yield-prediction model was of the form: $V=27.53 + -21.41\log DB + 0.023THT^2$ having $R^2$, SEE and bias values of 90.5%, 0.599 and 0.005, respectively. It is recommended for future predictions.

Keywords: Structural Diversity, Species Composition, Richness, Yield Prediction

1. Introduction

Forests form an integral part of life on earth, providing a range of benefits at local, national and global levels, covering approximately 30.6% of the world’s total land mass [1]. Tropical forested areas in Nigeria support high levels of biodiversity, but human interferences have played a destructive role in natural forest existence [2]. The conservation of plant diversity has received considerably less attention than the conservation of animals, perhaps because plants lack the popular appeal of many animal groups [3]. As a result, plant conservation is greatly under-resourced in comparison with fauna conservation [4]. Yet, plants are much more important. Animals can provide proteins, leather, fur and other products, but none of these are necessities for human survival and well-being, and would not survive without reasonable collections of trees [5].

Forests provide a large collection of ecosystem functions and services. They produce wood for pulp and timber, and contribute to the regulation of the carbon and water cycles [6]. Besides, they host a major part of the world’s biodiversity, and provide numerous cultural services and possibilities for recreation activities [6, 7]. Several decades of biodiversity research in grasslands suggest that plant diversity is a key driver of ecosystem functioning; it enhances primary productivity and increases resistance against herbivores [6].

Despite the immense benefits of forest and trees, they have disappeared at an alarming rate due to socio-economic changes such as infrastructural transformation, increasing commercialization of agricultural production, and high rate of urbanization, which result in large-scale destruction of the
natural forests. It is difficult to regenerate the existing forests and establishing new plantations by individuals, private bodies and government due to lack of proper knowledge about most species, and their silvicultural requirements [8]. High species diversity is considered as a desirable state, and that the knowledge of species diversity is particularly useful when one wishes to study the influence of biotic disturbances, or to know the state of succession and stability in the ecosystem [9, 10]. The diversity of life forms may be directly correlated to tree species diversity, or inversely correlated as in transitional forest-types characterized by diversity patterns at different geographic scales. Within the National Parks of Nigeria are some of the very few remaining natural ecosystems in the country, capable of enhancing ecological processes and life support system [11]. These parks are part of the efforts of the governments of Nigeria at conserving forests.

The environmental significance of the parks is mainly to establish an ecologically and geographically balanced network of protected areas under the control and jurisdiction of the Federal Government. They are expected to play a pivotal role in the conservation of biological diversities in the country [12]. In furtherance of these objectives, her management is to advise the Federal Government on the development and preservation including the financial requirements for the implementation of such policies. According to Reddy and Ugle [13], sustainable management techniques are required to maintain the biodiversity and productivity of tropical forest ecosystems, and this can only be possible through genuine information about the status and distribution of tree species, which form the frame for other life forms. Many of the once diverse natural forests have been lost to the plantation of exotic species and agricultural practices [7]. Consequently, there are severe ecological and environmental changes, reducing the stabilizing functions of the forest. Although, biodiversity is conventionally measured in terms of genetics, species and ecosystem diversity [14, 15], Nigeria’s rich biodiversity is highly influenced by its enormous anthropogenic forces and the floral diversity has however been poorly documented.

Although there are many univariate techniques for data analysis in statistics, none of them takes into account the effects of other variables [16]. In such a case regression models are being used. Generally, measuring breast height diameter is simpler, more accurate, less time consuming and cheaper than measuring tree height. Consequently, in forest inventories, diameter is measured for all the sampled trees, but height is measured only for a subsample of trees. From these measurements, statistical models are fitted to define the relation between these two variables [17, 18]. The most common procedure is to use an established height-diameter model to predict tree heights from field measurements of tree diameters [19, 20].

Modelling is a method of developing equation using linear prediction function, and unknown regression equation parameters are estimated from the data [21]. Regression analyses are used for data description (descriptive factors), prediction, parameters estimation and control [22]. Regression analysis is mainly used to predict variable values by the values of one or more independent variables [23]. Tree slenderness coefficient (TSC) has been used as the simplest empirical stability indicator for single tree or stand dimensions [24]. It is the ratio of tree total height to diameter at breast height (Dbh), with both variables measured in the same unit [25]. According to Wang et al. [26], the susceptibility of a stand to wind-induced damage is largely influenced by the TSC. The size of trees, their shape and structure influence mechanical stability [27]. Therefore, in order to predict the stability and susceptibility of a stand to damage, accurate information about TSC of a stand becomes very essential. Furthermore, tree crown condition is a good indicator of the health of a tree as it plays a key role in tree primary productivity provides habitats for myriads of wildlife [8].

Literature is replete with different tree growth models, which predict yields of forests and in recent time, the study on TSC is becoming significant in quantitative forestry following the susceptibility of many exotic species to natural phenomenon such as wind damage, breaking or uprooting of live tree as a result of high intensity of wind components [24]. Sustainable forest and forest resources management requires reliable estimates of growing stock. This is because such information guides forest managers in timber evaluation as well as in the allocation of forest areas for harvest [18, 28]. For timber production as well as other purposes, an estimate of growing stock is often expressed in terms of volume, which can be estimated from easily measurable tree dimensions such as diameter at breast height and tree height. The roles of models in tree volume estimation especially in tropical natural forest ecosystem cannot be overestimated [29]. Models are veritable tools for effective management of any forest stand. Models in forestry are tools for providing long-term decision-making, estimation of growing stock, timber valuation and allocation of forest areas for harvest. According to Avery and Burkhart [30], volume equations can be used to estimate the average content of standing trees of various sizes and species. Akindele and LeMay [31] reported that growing stock in forestry is usually expressed in terms of timber volume and the most common procedure of obtaining this is the use of allometric equations based on relationship between volume and variables such as diameter and height. The main reason for conducting forest inventory either in the plantations or natural forest ecosystem is to estimate timber volume of the plots installed in the entire stand.

Nigeria’s rich biodiversity is highly influenced by enormous anthropogenic forces and the floral diversity has however been poorly-documented, especially in most of the gazetted areas with a particular reference to Old Oyo National Park. Up till now, information on status of the trees in the park appears to be non-existent. For a sound forest management option, baseline data are crucial. The knowledge of species composition in the park remains inadequate despite the recognition of the immense relevance of trees. Structural information, which forms essential components of any sustainable forest management planning are yet unavailable. Little is known about the tree diversity of the
park. Hitherto, no tree growth prediction models have been developed for sustainable management of the forest in the park. The ever-increasing demands for forest goods and services has brought about intense pressure on the forest ecosystem and forest products, thereby leading to rapid disappearance of forest and loss of important species in natural habitats [7]. As a result of this pressure, many once-diverse natural ecosystems have been lost to plantation of exotic species and agricultural practices. Thus, it is important to have information about the natural forest as this may facilitate the formulation of sustainable forest management strategies and effective conservation work.

2. Methodology

2.1. Study Area

Old Oyo National Park (OONP) is located across northern Oyo State and southern Kwara State. The park is 251,200 ha of land on latitudes 8°15’ and 9°00’N, and longitude 3°35’ and 4°42’E (Figure 1). The location has inevitably placed the park at a vantage position of abundance land area as well as diverse wildlife and cultural/historical settings. Vegetation of the area has been classified as southern guinea savannah. There are three watersheds in Old Oyo National Park: River Ogun and its numerous tributaries, River Tessi and its tributaries and River Iwa and its tributaries. Annual rainfall in the park ranges between 900 mm and 1,500 mm, and mean annual temperature is between 12 and 37°C. The rainy season begins in April through September with peak between July and August. Temperature is highest in the dry season with the mean daily maximum, which is greatest during February and March with about 33.6°C with the lowest values of about 20°C during the peak of harmattan in December and January.

2.2. Sampling Procedure and Data Collection

Systematic line transect technique was adopted for sample plot location. Thirty transects of 1000 m long, evenly distributed over the five (5) camps of the OONP were laid at 1 km intervals. Four plots of 50 m × 50 m (0.25ha) were then laid alternately at 250 m intervals along each transects (Figure 2). A total of 120 sample plots (30 ha) were used for the study. The service of an experienced taxonomist was employed for tree species identification. In the study area, all living trees with Dbh ≥10 cm were identified by species, and measured. Tree growth parameters were measured using diameter tape, and Spiegel relaskop for measuring heights, diameter at the top and middle of the trees. All other important species below 10 cm in Dbh were recorded. All the
identified species were defined based on their IUCN status using IUCN Red List of Threatened species [32].

**Figure 2.** Sample plots' layout using systematic sampling technique

### 2.3. Data Analysis

The data on tree species were analyzed using, relative density, Shannon-Wiener diversity index, species richness, stand density and slenderness coefficient. Relative density (RD) for each of the tree species was determined using:

\[
RD(\%) = \left( \frac{n_i}{N} \right) \times 100
\]

(1)

Where; \(n_i\)=number of individual species, \(N\)=total number of species in the sampled area.

Tree species diversity was computed using Shannon-Wiener Diversity Index as:

\[
H' = -\sum_{i=1}^{S} P_i \ln(P_i)
\]

(2)

Where; \(S\)=total number of species in the community, \(P_i\)=proportion of \(S\) made up of the \(i\)th species, \(ln=\)natural logarithm.

Tree species richness was computed using Margalef’s index as:

\[
d = \frac{S}{\sqrt{N}}
\]

(3)

Where; \(S\)=the number of species encountered, \(N\)=the total number of individuals of all the tree species.

**Basal area**

Individual tree basal areas were computed using:

\[
BA = \frac{nDBh^2}{4}
\]

(4)

Where; \(DBh=\)diameter at breast height (m).

**Slenderness coefficient**

Tree slenderness coefficient (SC) for all measured trees was computed using:

\[
SC = \frac{\text{Total height}}{\text{DBh}}
\]

(5)

**Volume estimation**

Individual tree volumes were computed using Newton’s formula [33]:

\[
V = \frac{h}{25} (D_b + 4D_m + D_t)
\]

(6)

Where; \(D_b, D_m, D_t\)=tree diameters at the base, middle and top of merchantable height in cm, respectively, \(h=\)tree merchantable height (m). The plot volume was obtained by adding the volumes of all the trees in the plot, and the per-hectare volumes were obtained by multiplying the plot volumes by 4.

**Descriptive statistics**

Descriptive statistics such as mean and graphs were used to summarize the tree growth parameters and stand density.

### 2.4. Modelling

The following model functions were tried:

1. \(V=b_0 + b_1 SC + b_2 SQ^2 + b_3 \log DBh\)
2. \(V=b_0 + b_1 SQ^2 + b_2 SC^2 + b_3 DBh\)
3. \(V=b_0 + b_1 \log SQ + b_2 \log BA + b_3 SC\)
4. \(V=b_0 + b_1 \log SQ^2 + b_2 SC + b_3 BA\)
5. \(V=b_0 + b_1 \log BA + b_2 \log SC + b_3 MHT^2\)

Where; \(SQ=\)stem quality, \(MHT=\)merchantable height, \(b_0, b_1, b_2\) and \(b_3=\)regression parameters to be estimated. Models were adopted from Adeyemi and Ugo-Mbonu [24], and modified for this study.

### 2.5. Models Assessment

**F-test**

This was used to test the overall significance of the models. The critical values of F (i.e., F-tabulated) at \(\alpha =0.05\), were compared with the variance-ratio. Where the variance ratio (F- calculated) is greater than the critical values (F-tabulated), such equation is significant, and can therefore be accepted for volume prediction in the study area.

**Coefficient of Determination (R^2)**

This is a measure of the proportion of variation in the dependent variable that is being explained by the behaviour of the independent variable. For the model to be accepted, the \(R^2\) must be high (at least, \(\geq50\%\)).

\[
R^2 = 1 - \frac{\text{SSE}}{\text{SST}}
\]

(12)

Where; \(SSE=\)error sum of squares, \(SST=\)total sum of squares.

**Standard Error Estimate (SEE)**

\[
\text{SEE} = \sqrt{n-k (V_{obs}-V_{pred})^2} / n-k
\]

(13)

Where; \(V_{obs}=\)observed (computed) volume value, \(V_{pred}=\)predicted volume values, \(k=\)number of parameters in the fitted model, \(n=\)number of trees in the model-fitting dataset.
2.6. Models Validation

For model validation, data were divided into two sets (i.e. model-formulating set and model-validating set). The model-formulating set was used for developing the models and the usefulness of the models were validated using the model-validating set. The following statistics were then computed:

The student t-test

This was used to test for significant differences between the actual volume values and the model outputs (predicted values) of the various models developed.

Bias

This was used to examine the absolute differences between the computed volumes and the model outputs, as:

$$Bias = \sum \frac{V_i - V_o}{V_o}$$  (14)

Where; $V_o$=actual (observed) volume, $V_i$=predicted volume from the models.

3. Results

A total of 64 tree species, belonging to 23 families were identified in the area. Burkea africana was the most frequently-occurring species in the area, having RD of 9.3% (Table 1). The least-occurring species were Lecaniodiscus cupanioides, Ficus thonningii, Lannea schimperi, Maranthes agnis, Myrianthus arboreus, Securidaca longipedunculata, Securinega virosa, Sterculia setigera, Strychnos spinosa and Vitex doniana with less than one tree/ha. The species richness index for the area was 1.189 and the diversity index was 3.544. Burkea africana has the highest basal area/ha of 3.058 m²/ha, followed by Vitellaria paradoxa with 2.367 m²/ha. The least-occurring species in the area were Lecaniodiscus cupanioides (0.010 m²/ha), Ficus thonningii (0.047 m²/ha), Lannea schimperi (0.005 m²/ha), Maranthes agnis (0.035 m²/ha) and Vitex doniana (0.028 m²/ha). The Fabaceae was observed to be the most prevalent family (26.6%). Most of the species encountered have not been assessed against the IUCN-standard. Only one (Pterocarpus erinaceus) of the species is endangered according to the IUCN Red List. There were three vulnerable species (i.e. Afzelia africana, Maranthes agnis and Vitellaria paradoxa) present in the area. Five tree species (Azadirachta indica, Detarium microcarpum, Isoberlinia doka, Nauclea latifolia and Vitex doniana) classified as Least Concern were also encountered in the area. Table 2 presents individual tree and stand structural variables in the five ranges of the study area.

| Species                  | Family      | N/ha | BA/ha (m²) | RD% | IUCN Status  |
|--------------------------|-------------|------|------------|-----|--------------|
| Acacia polycantha        | Fabaceae    | 6    | 0.012      | 1.5 | Not assessed |
| Acacia seyal             | Fabaceae    | 4    | 0.021      | 0.9 | Not assessed |
| Afzelia africana         | Fabaceae    | 2    | 0.077      | 0.5 | Vulnerable   |
| Annona senegalensis      | Annonaceae  | 27   | 0.037      | 7.2 | Not assessed |
| Anogeissus leiocarpa     | Combretaceae| 26   | 2.348      | 6.9 | Not assessed |
| Azadirachta indica       | Meliaceae   | 2    | 0.368      | 0.4 | Least concern|
| Azanza garkeana          | Malvaceae   | 2    | 0.072      | 0.5 | Not assessed |
| Bombax costatum          | Malvaceae   | 2    | 0.019      | 0.4 | Not assessed |
| Bridelia ferruginea      | Euphorbiaceae| 4    | 0.284      | 1.1 | Not assessed |
| Burkea africana          | Fabaceae    | 35   | 3.058      | 9.3 | Not assessed |
| Combretum molle          | Combretaceae| 5    | 0.525      | 1.2 | Not assessed |
| Combretum nigricans      | Combretaceae| 4    | 0.259      | 1.1 | Not assessed |
| Crassopteryx febrifuga   | Rubiaceae   | 5    | 0.371      | 1.2 | Not assessed |
| Cussonia arborea         | Araliaceae  | 4    | 0.165      | 1.0 | Not assessed |
| Cussonia bateri          | Araliaceae  | 4    | 0.407      | 1.0 | Not assessed |
| Daniellia oliveri        | Fabaceae    | 17   | 1.966      | 4.4 | Not assessed |
| Detarium microcarpum     | Fabaceae    | 15   | 1.806      | 3.9 | Least concern|
| Detarium senegalense     | Fabaceae    | 2    | 0.083      | 0.5 | Not assessed |
| Distemonanthes benthamianus| Fabaceae  | 2    | 0.101      | 0.4 | Not assessed |
| Entada africana          | Fabaceae    | 2    | 0.125      | 0.5 | Not assessed |
| Ficus sp.                | Moraceae    | 5    | 0.192      | 1.1 | Not assessed |
| Ficus thonningii         | Moraceae    | 1    | 0.048      | 0.2 | Not assessed |
| Gardenia aquilla         | Rubiaceae   | 9    | 0.019      | 2.4 | Not assessed |
| Gardenia sokotensis      | Rubiaceae   | 7    | 0.008      | 1.8 | Not assessed |
| Grewia mollis            | Malvaceae   | 10   | 0.032      | 2.5 | Not assessed |
| Hymenocercis acida       | Phyllanthaceae| 15  | 1.509      | 4.0 | Not assessed |
| Isoberrilia doka         | Fabaceae    | 7    | 0.588      | 1.7 | Least concern|
| Lecaniodiscus cupanioides| Sapindaceae | 0.04 | 0.010      | 0.04| Not assessed |
| Lannea acida             | Anacardiaceae| 3   | 0.205      | 0.7 | Not assessed |
| Lannea schimperi         | Anacardiaceae| 0.04| 0.005      | 0.04| Not assessed |
| Lophira lanceolata       | Ochnaceae   | 8    | 0.649      | 2.1 | Not assessed |
| Maranthes polyandra      | Chrysobalanaceae| 3 | 1.138      | 0.8 | Not assessed |
| Maranthes agnis          | Chrysobalanaceae| 1 | 0.035      | 0.2 | Vulnerable   |
| Maytenus senegalensis    | Celastraceae| 5    | 0.211      | 1.2 | Not assessed |
| Monotes kerstenii        | Dipterocarpaceae| 7 | 0.576      | 1.8 | Not assessed |

Table 1. Tree species composition and their IUCN status in OONP.
The result of slenderness coefficient (SC) categorization for trees in the study area revealed that about 177 trees/ha had low slenderness coefficient (i.e. SC <70), constituting 76.2% of the trees encountered per hectare. This implies high stability, and low susceptibility to wind-induced damages or breakages. There were 30 trees/ha (13.1%) with moderate slenderness coefficient (i.e. SC: 70-80). Trees with high slenderness coefficient (SC >80) were 25/ha (10.7%), implying that very few of the trees/ha might be susceptible to wind-throw (Figure 3).

Parameter estimates of the models for predicting tree stem volume of species encountered in the study area are presented in Table 3. The intercepts (a) were far from 1 under models 1, 2, 3, 5 and 6 while the slope (b) were close to zero under models 1, 2, 3, 4, 6 and 7. Model 5 had the farthest value of slope from zero (8.765).

Figure 3. SC classes for tree species in the study area.

Table 2. Descriptive statistics for individual tree and stand growth variables.

| Species | Family | N/ha | BA/ha (m²) | RD% | IUCN Status |
|---------|--------|------|------------|-----|-------------|
| Myrianthus arboreus | Urticaceae | 1 | 0.719 | 0.2 | Not assessed |
| Nauclea latifolia | Rubiaceae | 5 | 0.282 | 1.3 | Least concern |
| Parinari maranthes | Chrysobalanaceae | 5 | 0.285 | 1.3 | Not assessed |
| Parinari polyandra | Chrysobalanaceae | 5 | 0.367 | 1.1 | Not assessed |
| Parkia biglobosa | Fabaceae | 3 | 0.117 | 0.8 | Not assessed |
| Pericopsis laxiflora | Fabaceae | 2 | 1.693 | 0.6 | Not assessed |
| Piliostigma thonningii | Fabaceae | 22 | 0.553 | 5.9 | Not assessed |
| Prosopis africana | Fabaceae | 7 | 1.186 | 1.9 | Not assessed |
| Pseudocedrela kotschyi | Meliaceae | 10 | 0.841 | 2.6 | Not assessed |
| Pterocarpus erinaceus | Fabaceae | 7 | 0.23 | 1.7 | Endangered |
| Securidaca longipedunculata | Polygalaceae | 1 | 0.074 | 0.2 | Not assessed |
| Securinea virosa | Euphorbiaceae | 1 | 0.041 | 0.3 | Not assessed |
| Sterculia setigera | Sterculiaceae | 1 | 0.141 | 0.1 | Not assessed |
| Stereospermum kunthianum | Bignoniaceae | 3 | 0.142 | 0.7 | Not assessed |
| Strychnos innocua | Loganiaceae | 3 | 0.036 | 0.8 | Not assessed |
| Strychnos spinosa | Loganiaceae | 1 | 0.905 | 0.2 | Not assessed |
| Terminalia glaucescens | Combretaceae | 11 | 1.962 | 3.0 | Not assessed |
| Terminalia micropetra | Combretaceae | 17 | 0.207 | 4.5 | Not assessed |
| Trichilia emetic | Meliaceae | 5 | 0.243 | 1.1 | Not assessed |
| Uapaca togoensis | Phyllanthaceae | 3 | 2.637 | 0.8 | Not assessed |
| Vitellaria paradoxa | Sapotaceae | 25 | 0.028 | 6.6 | Vulnerable |
| Vitex domiana | Verbenaceae | 1 | 0.074 | 0.1 | Least concern |

| Total | 378 | 28.89 | 100 |

N - number of trees; BA - basal area.
The model assessment results for species in the study area are presented in Table 4. All the models were good for prediction going by their $R^2$-values (at least, >50%) and significance of the models (P<0.05). All models except model 6 had SEE values of > 1 (Table 4). The best model was model 6 with $R^2$-value and SEE value of 90.5% and 0.599, respectively. This was followed by model 4 with $R^2$ and SEE values of 72.6% and 1.015, respectively. The least-suitable model was model 2 with $R^2$ and SEE values of 64.2% and 1.163, respectively.

### Table 3. Parameter estimates of the models.

| SN | Models | Parameter estimate | a   | b1  | b2  | b3  | b4 |
|----|--------|--------------------|-----|-----|-----|-----|----|
| 1  | $V=a + b_1 SC + b_2 SQ^2 + b_3 log Dbh$ | -5.382 | 0.184 | 0.374 | 1.114 |
| 2  | $V=a + b_1 SC^2 + b_2 Dbh$ | -2.297 | 0.036 | 0.003 | 0.094 |
| 3  | $V=a + b_1 log SQ + b_2 log BA + b_3 SC$ | 5.807 | 0.250 | 7.607 | 0.200 |
| 4  | $V=a + b_1 log SQ^2 + b_2 SC + b_3 BA$ | -1.007 | 0.349 | 0.136 | 9.615 |
| 5  | $V=a + b_1 log BA + b_2 log SC + b_3 MHT$ | -1.312 | 8.765 | 9.688 | -0.027 |
| 6  | $V=a + b_1 log DB + b_2 THT^2 + b_3 DB$ | 27.529 | -21.408 | 0.023 | 0.194 |
| 7  | $V=a + b_1 SQ^2 + b_2 THT + b_3 log BA$ | 2.645 | -0.006 | 0.501 | 3.862 |

The results of model validation are presented in Table 5. The result revealed that most of the developed models differed insignificantly in mean observed and predicted stem volumes. Models 1, 2 and 5 are not good models going by their significance in the observed and predicted volumes. However, models 3, 4, 6 and 7 were good for prediction since the predicted volume values were not significantly different from the measured (observed) values (P>0.05). The bias values of the four models were also very small (Table 5).

### Table 4. Model assessment results.

| SN | Model | F  | P   | $R^2$ | SEE |
|----|-------|----|-----|-------|-----|
| 1  | $V=-5.382 + 0.184 SC + 0.374 SQ^2 + 1.114 log Dbh$ | 1541.264 | 0.000 | 72.7 | 1.014 |
| 2  | $V=-2.297 + 0.036 SQ^2 + 0.003 SC^2 + 0.094 Dbh$ | 1037.865 | 0.000 | 64.2 | 1.163 |
| 3  | $V=-5.805 + 0.250 log SQ + 7.607 log BA + 0.200 SC$ | 1076.738 | 0.000 | 65.0 | 1.148 |
| 4  | $V=-1.017 + 0.349 log SQ^2 + 0.136 SC + 9.615 BA$ | 1538.196 | 0.000 | 72.6 | 1.015 |
| 5  | $V=-1.312 + 8.765 log BA + 9.688 log SC + 0.027 MHT$ | 1338.745 | 0.000 | 69.8 | 1.066 |
| 6  | $V=-27.529 + 21.408 log DB + 0.023 THT^2 + 0.194 DB$ | 5506.149 | 0.000 | 90.5 | 0.599 |
| 7  | $V=-2.645 + 0.006 SQ^2 + 0.501 THT + 3.862 log BA$ | 1404.505 | 0.000 | 70.8 | 1.049 |

### Table 5. Model validation results.

| SN | Models | t-test | $p$-value | Bias |
|----|--------|--------|-----------|------|
| 1  | $V=-5.382 + 0.184 SC + 0.374 SQ^2 + 1.114 log Dbh$ | 59.344 | 0.000 | 0.8423 |
| 2  | $V=-2.297 + 0.036 SQ^2 + 0.003 SC^2 + 0.094 Dbh$ | 48.981 | 0.000 | 0.8204 |
| 3  | $V=-5.805 + 0.250 log SQ + 7.607 log BA + 0.200 SC$ | 0.075 | 0.940 | 0.0013 |
| 4  | $V=-1.017 + 0.349 log SQ^2 + 0.136 SC + 9.615 BA$ | 0.032 | 0.974 | 0.0006 |
| 5  | $V=-1.312 + 8.765 log BA + 9.688 log SC + 0.027 MHT$ | 5.665 | 0.000 | 0.0618 |
| 6  | $V=-27.529 + 21.408 log DB + 0.023 THT^2 + 0.194 DB$ | 0.419 | 0.675 | 0.0050 |
| 7  | $V=-2.645 + 0.006 SQ^2 + 0.501 THT + 3.862 log BA$ | 0.107 | 0.915 | -0.0008 |

### 4. Discussion

Tree species of about 64 in 23 families are lower compared to the 125 species belonging to 36 families reported by Adeyemi et al. [7] for Okwango Forest in northern Cross River State and 102 tree species belonging to 35 families reported by Edet et al. [34] for Afik Mountain Wildlife Sanctuary, which have some similarities with the study area. However, the species encountered were more than 10 tree species in 8 families recorded by Ibe et al. [35] in Ohaji/Egbema Watershed in Imo State. Vulnerable tree species that were identified based on IUCN categories in the course of this study include *Afzelia africana*, *Maranthes agnesis* and *Vitellaria paradoxa*. While *Isobertlina doka*, *Azadirachta indica*, *Detarium microcarpum*, *Nauclea latifolia* and *Vitex doniana* are Least Concern, *Pterocarpus erinaceus* is endangered, and the other tree species were not yet assessed [32]. The reason for the poor establishments of some families in the area may be attributed to anthropogenic activities prevalent in the study area, which might have prevented successful regenerations due to over-exploitation of some mother trees. It is also possible that most of the seeds or other propagules needed for growth enhancement were harvested for human consumption, and this might hinder propagation of those species. This is in line with Wardle et al. [36] observations, that anthropogenic activities have great deleterious consequences on the abundance of tree species.

The overall mean basal area per hectare recorded in this study was higher than the values of 11.39 m$^2$ reported by Akinsanmi and Akintunde [37] and 1.1091 m$^2$ reported by
Ibe et al. [35], implying better structural diversity. The higher basal area may be due to the presence of adapted root architecture to absorb nutrients for growth. This corroborates the work of Parthasarathy [38], who noted that the adaptation of particular species to an environment may enhance their growth and establishment. However, this negates the influence that human use-pressure might have on any ecosystem, especially in developing world, where up to 70% of the population leave in rural areas, and are heavily dependent on forest resources, especially non-timber forest resources. The study area is characterized by abundance of trees with small Dbh. This is similar to the finding of Jimoh et al. [39]. The possibilities are that most trees in the large diameter classes were constantly exploited, often times, illegally. This also gives an impression of the structure diameter classes were constantly exploited, often times, illegally. This also gives an impression of the structure.

All the measured trees in the area had low slenderness coefficient, which implies high stability, and low susceptibility to wind-induced damage or breakage. Trees with high slenderness coefficient are more susceptible to breakage than those with low slenderness coefficients [24, 40]. The best volume prediction model is of the form: $V=27.529 + (-21.408\log_{10}DB) + 0.023THT^2 + 0.194DB$. This model had properties, which hitherto, had not been used or considered. Diameter at the base (DB) was rarely seen in use for modeling, but its efficacy in approximating stem volume prediction is really noteworthy. It is a parameter, which is very easy to measure with inexpensive instruments. The standard error of estimate is a good measure of overall predictive value of regression equations [9, 18, 31], and the value was really very small, just as the bias values, particularly for this prediction-suitable model. Generally, in this study, standard error of estimate (SEE) ranged between 0.599 and 1.163. Models with high index of fit such as $R^2$ are indicators of a suitable model with good fit [20, 24, 29].

5. Conclusion

This study has shown that OONP has moderately-high species diversity. However, use-pressure impacted negatively on the local species abundance. Although it appeared that conservation efforts in the study area are worthwhile, much needed to be done to ensure that over-exploitations of valuable trees species are halted. Families noted with dominant species in the area are Fabaceae and Combretaceae. However, most tree species encountered in the area are not yet assessed against IUCN-standard, and only very few species can be said to be abundant in the area due to use-pressure. The overall mean Dbh (16.76 cm) in the area was lesser than the national minimum (48 cm) as stipulated by logging policy of Nigeria. This means, there should not be exploitation in the area in the first place. Although most of the trees measured had low to medium slenderness coefficients, which implies that most of the trees were of good vigour and could withstand wind-throw, this may have resulted from the fact that tall trees were already selectively logged, and illegally, since the area is under protection by law. Hence, responsible agency should be more sensitively-intolerant to all forms of logging in the area. The park management should mostly concentrate on blocking known leakages, and make all culprits to face full wrath of the law, otherwise the area might soon be a history. There should be frequent tree species monitoring involving biodiversity and structural assessments for possible changes, either positive or otherwise. Tree growth parameters, especially tree diameter at the base and height proved to be good predictors of tree volume. The best model developed was of the form: $V=27.529 + (-21.408\log_{10}DB) + 0.023THT^2 + 0.194DB$ with $R^2$, SEE and bias values of 90.5%, 0.599 and 0.0050%, respectively. It is therefore recommended for predictions in the study area.

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