Adaptive Evolution of *Sphingobium hydrophobicum* C1<sup>T</sup> in Electronic Waste Contaminated River Sediment

**Da Song**<sup>1,2</sup>, **Xingjuan Chen**<sup>2</sup>, **Meiying Xu**<sup>2,*</sup>, **Rong Hai**<sup>3</sup>, **Aifen Zhou**<sup>4</sup>, **Renmao Tian**<sup>4</sup>, **Joy D. Van Nostrand**<sup>4</sup>, **Megan L. Kempher**<sup>4</sup>, **Jun Guo**<sup>2</sup>, **Guoping Sun**<sup>2</sup> and **Jizhong Zhou**<sup>4</sup>

1 School of Biology and Biological Engineering, South China University of Technology, Guangzhou, China, 2 State Key Laboratory of Applied Microbiology Southern China, Guangdong Provincial Key Laboratory of Microbial Culture Collection and Application, Guangdong Institute of Microbiology, Guangdong Academy of Sciences, Guangzhou, China, 3 Department of Plant Pathology and Microbiology, University of California, Riverside, Riverside, CA, United States, 4 Institute for Environmental Genomics, Department of Microbiology and Plant Biology, University of Oklahoma, Norman, OK, United States

Electronic waste (e-waste) has caused a severe worldwide pollution problem. Despite increasing isolation of degradative microorganisms from e-waste contaminated environments, the mechanisms underlying their adaptive evolution in such habitats remain unclear. Sphingomonads generally have xenobiotic-degrading ability and may play important roles in bioremediation. *Sphingobium hydrophobicum* C1<sup>T</sup>, characterized with superior cell surface hydrophobicity, was recently isolated from e-waste contaminated river sediment. To dissect the mechanisms driving its adaptive evolution, we evaluated its stress resistance, sequenced its genome and performed comparative genomic analysis with 19 other *Sphingobium* strains. Strain C1<sup>T</sup> can feed on several kinds of e-waste-derived xenobiotics, exhibits a great resistance to heavy metals and possesses a high colonization ability. It harbors abundant genes involved in environmental adaptation, some of which are intrinsic prior to experiencing e-waste contamination. The extensive genomic variations between strain C1<sup>T</sup> and other *Sphingobium* strains, numerous C1<sup>T</sup>-unique genes, massive mobile elements and frequent genome rearrangements reflect a high genome plasticity. Positive selection, gene duplication, and especially horizontal gene transfer drive the adaptive evolution of strain C1<sup>T</sup>. Moreover, presence of type IV secretion systems may allow strain C1<sup>T</sup> to be a source of beneficial genes for surrounding microorganisms. This study provides new insights into the adaptive evolution of sphingomonads, and potentially guides bioremediation strategies.

**Keywords:** *Sphingobium*, electronic waste (e-waste), xenobiotic degradation, heavy metal resistance, comparative genomics, genome plasticity, adaptive evolution

**INTRODUCTION**

Electronic waste (e-waste) has become one of the most rapidly growing pollutants worldwide. It contains various toxic compositions, such as heavy metals, polychlorinated biphenyls (PCBs), polybrominated diphenyl ethers (PBDEs), phthalate esters (PAEs), and polycyclic aromatic hydrocarbons (PAHs). Severe contamination has been detected around e-waste disposal areas, especially in river sediment due to the strong adsorption of pollutants and slow natural attenuation (Leung et al., 2006; Zhang K. et al., 2012; Ma et al., 2013).
Environmental stress plays a critical role in the evolution of organisms, and extreme stress may lead to extinction, evolutionary changes and speciation (Nevo, 2011). The complex combined pollution resulting from e-waste has significantly altered soil and sediment microbiomes (Liu et al., 2015, 2018). Understanding the microbial adaptive evolution mechanisms for such complexly contaminated environments will contribute to modeling function-oriented evolution, obtaining beneficial biological parts and designing bioremediation strategies. Some microorganisms capable of degrading typical e-waste pollutants have been isolated from e-waste contaminated environments (Hu et al., 2015; Tang et al., 2016). However, little is known about their adaptive evolution processes.

Sphingomonads (Sphingomonas in a broader sense including Sphingobium) are often isolated from environments contaminated by organic pollutants and generally have the capability to degrade a variety of recalcitrant organic compounds, such as PAHs and halogenated aromatics, suggesting that they adapt well to contaminated environments and play important roles in bioremediation (Stolz, 2009). They mainly employ abundant oxygenases to degrade these xenobiotics (Stolz, 2009). They can synthesize carotenoids which protect cells from reactive oxygen species produced during degradation process (Liu et al., 2012). Their general physiological characteristics, including special outer membrane component glycosphingolipids, biosurfactant excretion, biofilm formation and chemotaxis, help them access to nutrients or pollutants (Cunliffe and Kertesz, 2006; Coppotelli et al., 2010). Comparative genomic analysis of sphingomonads have revealed diversity in the genomic organizations and genetic characteristics involved in pollutant degradation (Aylward et al., 2013; Verma et al., 2014; Wang et al., 2018), marine adaptation (Gan et al., 2013), nitrate respiration (García-Romero et al., 2016), etc. However, to the best of our knowledge, only two sphingomonads, Sphingobium fuliginis HC3 and Sphingomonas sp. MXB8, isolated from e-waste contaminated environments have been reported, and no genomic information is available. They are both from soil samples. Strain HC3 can degrade biphenyl and PCBs without dead-end intermediates accumulation (Hu et al., 2015), while strain MXB8 shows an excellent potential for the bioleaching of Ag from e-waste (Díaz-Martínez et al., 2019).

Sphingobium hydrophobicum C1T was recently isolated from the sediment of the Lianjiang River in Guiyu, China, which had been contaminated for more than two decades by wastewater discharged from e-waste disposal (Chen et al., 2016). The sediment contains high concentration of heavy metals (Cu, 528 mg/kg; Zn, 249 mg/kg; Ni, 120 mg/kg; etc.), PAHs (3034 µg/kg), PBDEs (9054 µg/kg), PCBs (743 µg/kg) (Supplementary Table S1). Strain C1T is the most hydrophobic s h ingomonad ever known. Its high cell surface hydrophobicity (CSH), which may enhance colonization ability and adsorption of hydrophobic nutrients, is greatly attributed to the increased expression of certain outer membrane proteins (Chen et al., 2017). This study aims to address (i) what metabolic potential and genes enable strain C1T to adapt to the stress from e-waste contaminated sediment? (ii) what are the main mechanisms driving its adaptive evolution? We evaluated its capability to cope with the main environmental stressors, sequenced its genome and performed comparative genomic analysis with 19 other Sphingobium strains isolated from different environments. The results demonstrated advantageous physiological characteristics, beneficial genetic elements and high genome plasticity of strain C1T, highlighted the important role of horizontal gene transfer (HGT) in its adaptive evolution, and suggested its role of beneficial-gene contributor.

**MATERIALS AND METHODS**

**Strains and Culture Conditions**

*Sphingobium hydrophobicum* C1T (= CCTCC AB 2015198 = KCTC 42740) was isolated from e-waste contaminated sediment in Guiyu, China (Chen et al., 2016). Strain C2, a hydrophilic variant of strain C1T, was obtained by cell pass for ~100 generations in LB medium (Chen et al., 2017). Strains C3 and C4 were isolated from the first-round subculture (within 10 generations) of strains C1T and C2 in LB medium, respectively. *Sphingobium xenophagum* NBRC 107872 was purchased from China General Microbiological Culture Collection Center. *E. coli* ATCC 25922 was deposited in our laboratory. *Sphingobium* strains were cultivated aerobically in LB medium or mineral salt medium (Chen et al., 2016) at 30°C. *E. coli* strain was cultivated aerobically in LB medium at 37°C.

**Xenobiotics Degradation Assay**

Strain C1T was cultivated in the mineral salt buffer with different xenobiotics (~500 mg/L) as the sole carbon source. The tested xenobiotics included dimethyl phthalate, dibutyl phthalate, biphenyl, diphenyl ether, chlorobenzene, bromobenzene, etc. (Table 1). The growth of strain C1T, which was determined by naked-eye observation and protein quantification, indirectly

| Xenobiotics         | Growth |
|---------------------|--------|
| Dimethyl phthalate  | +      |
| Dibutyl phthalate   | +      |
| Biphenyl            | +      |
| p-Xylene            | +      |
| Ethybenzene         | +      |
| Phenol              | +      |
| Diphenyl ether      | +      |
| Chlorobenzene       | +      |
| Bromobenzene        | +      |
| 1,2-Dichlorobenzene | –      |
| Phenylamine         | –      |
| Benzene             | –      |
| Phenanthrene        | –      |
| Pyrene              | –      |
| n-Hexadecane        | –      |

*+, Megascopic cell aggregates appear and/or the medium becomes turbid within 7 days. --*, no obvious growth within 7 days.

---

**TABLE 1** | The xenobiotic-degrading ability of strain C1T.
reflected its degradation ability. Megascopic cell aggregates would appear and/or the medium would become turbid if strain C1T could grow. The protein was quantified by the Coomassie brilliant blue method as described previously (Fang et al., 2015). All cultures were prepared in triplicate.

**Heavy Metal Resistance Assay**

The bacterial resistance to Cr(VI), Cu²⁺, Zn²⁺, Hg²⁺, Cd²⁺, Ni²⁺, and Pb²⁺ was determined by observing the growth in LB medium containing the corresponding compounds [K₂Cr₂O₇, CuCl₂, ZnCl₂, HgCl₂, CdCl₂, NiCl₂, Pb(NO₃)₂]. The number and size of bacterial colonies were recorded for solid culture, while the optical density of cultures at 600 nm was measured for liquid culture. All assays were performed in triplicate.

**CSH Assay**

The CSH of cells in the late-exponential growth phase was measured using the microbial adhesion to hydrocarbon (MATH) protocol as described previously (Chen et al., 2017).

**Adhesion Assay**

The bacterial adhesion capacity to kaolinite was evaluated as described previously (Yee et al., 2000) with slight modifications. Briefly, cells were harvested after overnight growth in LB medium and washed three times with 1 mM KNO₃. Weighted cells were suspended in 30 mL of 1 mM KNO₃, placed in contact with 4 g/L kaolinite, and equilibrated by shaking at 150 rpm for 1.5 h. Separation of kaolinite-adsorbed and free cells was performed by horizontal centrifugation (5000 rpm, 15 min) with 60 wt% sucrose. The free cells floating on top of the sucrose layer were harvested, measured using a spectrophotometer (600 nm), and then converted into weight. The adhesion intensity was determined by subtracting the weight of free cells from the initial weight. All assays were performed in triplicate.

**Biofilm Formation Assay**

Biofilm formation ability was determined as described previously (Grzegorz et al., 2016) with slight modifications. Overnight cultures were diluted 100 times with LB medium and transferred into 96-well plates (150 µL/well). After 12, 24, 36, and 48 h static incubation, the biofilms in wells were washed three times with 0.9% NaCl and quantified by crystal violet staining. All assays were performed in sextuplicate.

**Replicons Detection**

The replicons in strain C1T and its variants (C2, C3, C4) were detected by in-gel cell lysis and pulse field gel electrophoresis. Salmonella enterica serovar Braenderup H9812 digested with XbaI was used as a size marker.

**Genome Sequencing**

The genomes of strains C1T and C2 were sequenced on PacBio RS II sequencing platform (Pacific Biosciences, United States).

**Data Analysis**

**Statistical Analysis**

Multiple comparisons of means were performed using SPSS (v20.0) with one-way ANOVA. Bonferroni test was employed for equal variances assumed, while Tamhane's T2 was employed for unequal variances.

**Genome Assembly**

*De novo* assembly of reads was performed by SMRT analysis pipeline v2.3.0 (Chin et al., 2013).

**Gene Prediction and Functional Annotation**

Open reading frames (ORFs) were identified by Prodigal (Hyatt et al., 2010). The rRNA, tRNA and other ncRNA were predicted using RNAmmer (Lagesen et al., 2007), ARAGORN (Laslett and Canback, 2004) and Infernal conjunction with Rfam (Nawrocki et al., 2009). The function of ORFs was annotated with EggNOG (Huertacepas et al., 2017), KEGG (Kanehisa et al., 2015), Swissprot and Nr. The β-barrel outer membrane proteins were predicted by PRED-TMBB² (Tsirigos et al., 2016). Two-component regulatory systems were predicted by P2RP² (Barakat et al., 2013).

**Phylogenetic Analyses**

The phylogenetic tree based on 16S rRNA gene sequences was constructed using Mega 6.0 with ClustalW alignment method and Neighbor Joining algorithm (Tamura et al., 2013). Another phylogenetic tree based on shared gene families was constructed by CMG-biotools (Vesth et al., 2013). The average nucleotide identity (ANI) values were calculated using the JSpeciesWS server¹ (Richtel et al., 2016).

**Prediction of Mobile Genetic Elements**

Prophages were predicted by PHASTER¹ (Arndt et al., 2016). Genomic islands (GIs) were predicted by IslandViewer4² using the method Sigi-HMM and IslandPath-DIMOB both based on sequence composition (Bertelli, 2017). Insertion sequences (ISs) were detected by ISsaga³ (Varani et al., 2011). All the analyses were performed with default parameters.

**Genomic Comparison Approaches**

The C1T genome was compared with the 19 Sphingobium genomes available (accession numbers in [Supplementary Table S2](https://www-is.biotoul.fr)) in the NCBI Genome Database. The core- and pan-genome analysis of these *Sphingobium* strains was performed using CMG-biotools with a threshold of 50% identity and 50% coverage for protein sequences (Vesth et al., 2013). The comparisons of genome/gene sequences were performed using BRIG (Alikhan et al., 2011), Mauve (Darling et al., 2004), Mummer (Kurtz et al., 2004), and MCScanX (Wang et al., 2012).

---

¹[http://www.compgen.org/tools/PRED-TMBB2](http://www.compgen.org/tools/PRED-TMBB2)
²[http://www.p2rp.org/](http://www.p2rp.org/)
³[http://jspecies.ribohost.com/jspeciesws/#Home](http://jspecies.ribohost.com/jspeciesws/#Home)
⁴[http://phaster.ca/](http://phaster.ca/)
⁵[http://www.pathogenomics.sfu.ca/islandviewer/browse/](http://www.pathogenomics.sfu.ca/islandviewer/browse/)
⁶[https://www.is.biotoul.fr](https://www.is.biotoul.fr)
Test for Positive Selection
The identification of orthologs between strain C1⁹T and *S. xenophagum* strains was based on two-way best Blastp match (coverage ≥ 70%, identity ≥ 80%). Orthologs were further aligned by ParaAT2.0 (Zhang Z. et al., 2012). Then Ka/Ks (non-synonymous to synonymous substitution rate ratio) analysis for orthologs was performed using KaKs_Calculator Toolbox 2.0 with the maximum-likelihood method GY (Wang et al., 2010). The gene was judged to be under positive selection if Ka/Ks > 1 and P-value (Fisher) < 0.05.

RESULTS

Stress Resistance of Strain C1⁹T
For microbial survival and growth, the main stressors from e-waste contaminated river sediment may include toxic xenobiotics, heavy metals, current scour and frequent environmental fluctuations. Therefore, strain C1⁹T was expected to degrade several kinds of e-waste-derived xenobiotics, exhibit a great resistance to heavy metals and possess a high colonization ability.

First, xenobiotic-degrading ability of strain C1⁹T was tested, and result showed that strain C1⁹T could utilize various common xenobiotics as the sole carbon source for growth, such as PAEs, biphenyl, diphenyl ether, bromobenzene and chlorobenzene (Table 1 and Supplementary Figure S1).

Heavy metal resistance of strain C1⁹T, *S. xenophagum* NBRC 107872 and *E. coli* ATCC 25922 was evaluated and compared. Strain NBRC 107872 is one of the most closely related strains. Strain ATCC 25922, a well-characterized Gram-negative strain, is widely used as a control for various laboratory experiments. As shown in Figure 1, strain C1⁹T could grow in LB medium containing Cu²⁺ (3 mM), Cr(VI) (0.3 mM), Ni²⁺ (1.5 mM), Pb²⁺ (3 mM), Cd²⁺ (1 mM), Hg²⁺ (0.25 mM) and Zn²⁺ (6 mM), respectively. It was more resistant to most of the tested heavy metals than strain NBRC 107872, and had a greater tolerance to Zn²⁺, Cd²⁺, and Hg²⁺ than strain ATCC 25922 (Figure 1 and Table 2).
### TABLE 2 | Heavy metal resistance of strains C1T, NBRC 107872 and ATCC 25922.

| Concentration (mM) | C1T | NBRC 107872 | ATCC 25922 |
|--------------------|-----|-------------|-------------|
| Cr(VI) 0.2         | ±   | –           | ±           |
| 0.3                | ±/– | –           | ±/–         |
| 0.4                | –   | –           | –           |
| Cu²⁺ 2             | +   | +           | +           |
| 3                  | ±/– | ±           | ±/–         |
| 4                  | –   | –           | –           |
| Zn²⁺ 2             | +   | –           | –           |
| 4                  | ±/– | –           | ±/–         |
| Hg²⁺ 0.1           | +   | –           | –           |
| 0.25               | ±   | –           | –           |
| Ni²⁺ 0.5           | ±   | –           | ±           |
| 1                  | ±   | –           | –           |
| Cd²⁺ 0.5           | ±   | –           | ±/–         |
| 1                  | ±   | –           | –           |
| 2                  | –   | –           | –           |
| Pb²⁺ 2             | +   | +           | +           |
| 3                  | +   | –           | ±           |
| 4                  | –   | –           | –           |

*+, the growth rate and number of bacterial colonies are similar to those in control group (growth in LB solid medium without heavy metals); ±, slower growth; +/–, slower growth and fewer bacterial colonies; –, no bacterial colonies.

In order to measure colonization ability of strain C1T, the CSH, adhesion capacity and biofilm-forming ability of strains C1T, C2, NBRC 107872 and ATCC 25922 were tested and compared. As shown in Figure 2, strain C1T displayed the highest CSH, strongest adhesion capacity to kaolinite (a main soil/sediment component) and highest biofilm-forming ability, suggesting that it might be able to colonize efficiently in sediment.

**General Features of the C1T Genome**

The complete C1T genome is 4.6 Mb and consists of two chromosomes and five large plasmids. General features are summarized in Table 3. The two largest replicons harbor 1 or 2 copies of rRNA gene operons, so are defined as chromosomes. The existence of a plasmid replication initiator in Chromosome 2 and its megaplasmid-like size suggest that it may originate from a plasmid followed by the uptake of some essential genes including rRNA genes. In total, 4506 ORFs are predicted, approximately 30% of which encode hypothetical proteins.

**Genomic Differences Between Strain C1T and Other Sphingobium Strains**

In order to find the genomic characteristics of strain C1T involved in adaptive evolution, the C1T genome was analyzed and compared with the genomes of 19 other *Sphingobium* strains isolated from different habitats (Supplementary Table S2). Based on 16S rRNA genes, shared gene families and ANI values, strain C1T is most closely related to *S. xenophagum* QYY and NBRC.
TABLE 3 | General features of the C1T genome.

| Size (Kb) | GC% | ORFs | rRNA | tRNA | tmRNA | miscRNA |
|----------|-----|------|------|------|-------|---------|
| Chromosome 1 | 3059.0 | 63.6 | 2979 | 3 | 52 | 1 | 6 |
| Chromosome 2 | 711.6 | 61.96 | 658 | 6 | 6 | 0 | 2 |
| Plasmid 1 | 270.5 | 63.89 | 278 | 0 | 0 | 0 | 1 |
| Plasmid 2 | 247.5 | 62.5 | 265 | 0 | 0 | 0 | 0 |
| Plasmid 3 | 171.6 | 62.37 | 185 | 0 | 0 | 0 | 1 |
| Plasmid 4 | 79.9 | 62.91 | 99 | 0 | 0 | 0 | 0 |
| Plasmid 5 | 62.6 | 59.26 | 62 | 0 | 0 | 0 | 0 |
| Total | 4802.6 | 63.19 | 4506 | 9 | 58 | 1 | 10 |
FIGURE 3 | Circular representation of the C1T genome and comparison results with other Sphingobium genomes. From inner to outer ring: (1) Seven replicons of strain C1T, (2) GC content, (3) GC skew, (4–21) Genome comparison (Blastn) results between strain C1T and other Sphingobium strains listed at the bottom, (22) Prophages predicted by PHASTER, (23) GIs predicted by IslandViewer4.
7 contains genes involved in type IV secretory system and heavy metal resistance; GI 21 encodes a complete peptide/nickel transport system and a haloacid dehalogenase. Interestingly, the plasmid sequences are found to have high similarity with plasmids of distant sphingomonads, and even different regions in a plasmid of strain C1T are similar to plasmids from different sources (Supplementary Figure S7).

For instance, the middle part of Plasmid 4 is similar to Plasmid unnamed1 of *Sphingomonas* sp. NIC1, and the both sides are similar to Plasmid of *Sphingopyxis* sp. 113P3. It suggests that plasmid transfer and rearrangement happen among sphingomonads. Additionally, the C1T genome harbors 5 gene clusters encoding three categories of type IV secretion systems, 1 *trb*, 2 *tra* and 2 *vir*, which are responsible...
for conjugation process of plasmids belonging to different incompatibility groups.

The genes and metabolic potential involved in coping with the main stressors were further analyzed.

**Xenobiotic Degradation**

A total of 8 monooxygenase genes and 19 dioxygenase genes responsible for cleaving the ring of aromatic/cyclic compounds were annotated in the C1\( ^T \) genome (Supplementary Table S5). They are scattered on Chromosome 1 (5), Chromosome 2 (15) and Plasmid 2 (7). Most have homologous sequences with high identity in strains QYY and NBRC 107872 (Figure 6A). Notably, the genes encoding alpha and beta subunits of biphenyl dioxygenase on Plasmid 2 are not present in the 19 other Sphingobium genomes. Instead, they are closely related to those of strains from Actinobacteria and next to an IS, suggesting that they were acquired via cross-phylum HGT. Two oxygenase genes (homogentisate 1,2-dioxygenase and camphor 5-monooxygenase) are found to be under positive selection (Supplementary Table S5).

The C1\( ^T \) genome harbors complete degradation gene clusters for catechol, protocatechuate and PAEs. The catechol degradation gene cluster is located on GI 10 flanked by ISs, without homologous sequences in strains NBRC 107872 and QYY, suggesting that this gene cluster was acquired via HGT. In contrast, the protocatechuate degradation gene cluster on Chromosome 2 has homologous sequences in strains NBRC 107872 and QYY with almost identical genome positions, suggesting that this gene cluster existed before these strains diverged.

The PAEs degradation gene cluster on Plasmid 2 comprises three regions separated by ISs (Figure 7). The first and third regions contain homologous genes for transforming phthalate to protocatechuate, suggesting a GD event. The
FIGURE 6 | Homology search of related protein sequences of strain C1T against the 19 other Sphingobium genomes. These proteins are listed in Supplementary Table S5. Each ring represents the Blastx comparison result against a genome. The presence/absence of homologous gene is visually presented, and the gradation of color indicates sequence identity. (A) Oxygenases and multidrug efflux pumps. (B) Heavy metal resistance proteins.
The PAEs degradation gene clusters of strains C1T, QYY and DS20.

(A) Organizations of these gene clusters. The purple blocks represent ISs. Other blocks contain genes responsible for corresponding metabolic processes.

(B) Homologous regions between these gene clusters detected by Mauve. The image scales of (A,B) are the same.

**FIGURE 7** The PAEs degradation gene clusters of strains C1T, QYY and DS20. (A) Organizations of these gene clusters. The purple blocks represent ISs. Other blocks contain genes responsible for corresponding metabolic processes. (B) Homologous regions between these gene clusters detected by Mauve. The image scales of (A,B) are the same.

The second region contains genes for degrading protocatechuate. Among the 19 other Sphingobium strains, only strains QYY and DS20 have homologous genes for transforming phthalate to protocatechuate. However, the genetic organizations are different among the three strains (Figure 7). These results suggest that the gene cluster may be derived from HGT initially and then undergo multiple translocations.

In addition, strain C1T has 4 sets of multidrug efflux pump genes, 3 *acrAB* and 1 *emrAB* (Supplementary Table S5), which may confer resistance to toxic xenobiotics. One *acrAB* is located on Chromosome 2, and the other three sets on Chromosome 1. Most *Sphingobium* strains have homologous genes, suggesting that these genes are intrinsic.

**Heavy Metal Resistance**

Strain C1T harbors 34 resistance genes for various heavy metals, such as chromate, copper, and mercury (Supplementary Table S5). Based on function annotation, toxic heavy metals could be either transformed into less toxic forms, or pumped out via efflux transporters. Among these genes, 11 are on Chromosome 1 (7 on GI 7), 6 on Chromosome 2, 5 on Plasmid 2 and 11 on Plasmid 4. Some resistance genes are multiple, e.g., 4 chromate
transporter genes (chrA), 3 cobalt-zinc-cadmium efflux gene clusters (czcABC) and 2 copper resistance gene clusters. The two chrA genes on Plasmid 1 and Plasmid 2 have identical sequences, and are flanked by similar sequences, suggesting a large-fragment GD event. However, other genes/gene clusters share low to median amino acid identity (<80%), and their locations in the genome have no correlation, indicating that they may not have directly originated from GD events.

The resistance genes on chromosomes (except those on GI 7) are closely related to strain QYY, while the others are closely related to distant sphingomonads. Especially the resistance genes on Plasmid 4 are closely related to strain C1 T sp. 113P3. Many resistance genes of strain C1 T have identical sequences, and their organizations are different (Supplementary Figure S9). The FixL-FixJ system, which constitutes an oxygen-sensitive switch for regulating genes involved in nitrogen fixation and/or microaerobic respiration, has two gene copies on Plasmids 1 and 3, respectively. No homolog is found in the 18 other Sphingobium genomes (except NBRC 16415). These results indicate that the two-component regulatory systems DctB-DctD and FixJ-FixL may be acquired via HGT.

**DISCUSSION**

Microorganisms survive and thrive in various natural or artificial environments and play important roles in geochemical cycling and environmental remediation. Sphingomonads are often studied due to their metabolic versatility, especially xenobiotic-degrading ability, but these studies often overlook other phenotypic and genetic characteristics involved in adaptation to contaminated environments (Stolz, 2009; Aylward et al., 2013; Tabata et al., 2016). In this study, we focused on the important physiological characteristics of S. hydrophobicum C1 T, a sphingomonad type strain isolated from e-waste contaminated river sediment, and employed comparative genomic analysis to uncover its adaptive evolution mechanisms.

The first question addressed is “what metabolic potential and genes enable strain C1 T to adapt to the environmental stress.” The stressors from e-waste contaminated river sediment generally include a variety of toxic xenobiotics and heavy metals, current and frequent environmental fluctuations. Strain C1 T could feed on several kinds of e-waste-derived xenobiotics, exhibited a great resistance to some heavy metals, and possessed a high colonization ability. As other sphingomonads, various oxygenase genes are responsible for its xenobiotic degradation potential (Aylward et al., 2013). The specific biphenoxy dioxygenase genes and PAEs degradation genes are involved in the degradation of e-waste-derived xenobiotics (Leung et al., 2006; Ma et al., 2013). Catechol and protocatechuate degradation gene clusters provide common downstream degradation pathways for aromatic compounds (Fuchs et al., 2011). Its heavy metal resistance is mainly attributed to multiple efflux pump genes, such as cobalt-zinc-cadmium efflux gene clusters (czcABC). Moreover, Hg^2+ and As(III) can be transformed to less toxic forms by MerA and ArsH, respectively (Silver, 1996; Yang and Rosen, 2016). The genes for exopolysaccharide synthesis may contribute to colonization ability, heavy metal resistance and removal (Periasamy et al., 2015; Gupta and Diwan, 2017). Abundant two-component regulatory systems allow it to respond to frequent environmental fluctuations. Among them, specific system FixL-FixJ may help it adapt to oxygen-deficient condition of sediment (Rodgers and Lukat-Rodgers, 2005). In brief, its genome harbors abundant genes involved in coping with the stressors (Figure 8A).
FIGURE 8 | The adaptive evolution of strain C1T in e-waste contaminated river sediment. (A) A brief metabolic reconstruction of strain C1T. The physiological processes and proteins involved in xenobiotic degradation, heavy metal resistance, environmental response and conjugational transfer are shown. (B) A conceptual model of the adaptive evolution process. Positive selection, gene duplication (GD), and especially horizontal gene transfer (HGT) drive the adaptive evolution of strain C1T. After rounds of selection, strain C1T obtained a high genome plasticity and advantageous phenotype.
The second question is “what are the main mechanisms driving the adaptive evolution of strain C1T.” Various microbial evolutionary mechanisms, including positive selection, GD and HGT, can result in a diverse array of metabolic potential, which in turn contributes to the rapid adaptation of microorganisms to environmental stress (Petersen et al., 2007; Andersson and Hughes, 2009; Hemme et al., 2016). However, the relative importance of these processes remains elusive. The genes involved in direct dynamic interactions with the environment are likely under positive selection (Petersen et al., 2007). GD occurs much more frequently than spontaneous point mutation (Andersson and Hughes, 2009). HGT may account for 1.6 ~32.6% of the genes in each individual genome (Koonin et al., 2001). Previous comparison of 26 sphingomonad genomes revealed that selfish genetic elements might be prominent forces shaping genomes, and that megaplasmids, prophages, transposons, and frequent genome rearrangements were prevalent features in this group (Aylward et al., 2013). Multiple degradation genes are usually scattered in the genome and flanked by insertion elements, which may allow quicker adaptation to xenobiotics than other bacteria by forming new composite degradation pathways (Stolz, 2009). In this study, explicit evidence is provided for high plasticity of the C1T genome, i.e., extensive genomic variations between strain C1T and other Sphingobium strains, numerous C1T-unique genes, massive mobile elements and frequent genome rearrangements. Numerous genes involved in coping with the stressors originate from HGT, especially HGT among sphingomonads, e.g., the biphenyl dioxygenase genes on Plasmid 2 and the heavy metal resistance genes on GI 7 and plasmid 4. Some genes under positive selection and GD events are detected, but a few of them are directly involved in xenobiotic degradation and heavy metal resistance. Therefore, HGT may be dominant during the evolution of xenobiotic degradation potential and heavy metal resistance.

In addition, genome comparison between strain C1T and its two most closely related strains QYY and NBRC 107872 provides clues about the origins of some genome elements. The history of e-waste recycling in the sediment sampling site Guiyu can be traced back to the late 1980s with a large increase in 1995 (Leung et al., 2006). Strains C1T, NBRC 107872 and QYY were isolated from different environments (Supplementary Table S2) before/in 2014 (Chen et al., 2016), 1986 (Nörtemann et al., 1986), and 2005 (Qu et al., 2006) respectively. Therefore, the genes of strain C1T, which have orthologs in strain NBRC 107872, are speculated to be intrinsic genes appearing prior to the e-waste contamination. These genes include the protocatechuate degradation gene cluster on Chromosome 2, multidrug efflux pump genes, most of the genes encoding two-component regulatory systems, etc. Likewise, the genes having orthologs in strain QYY can be speculated to appear before/during the early phase of e-waste contamination, e.g., some heavy metal resistance genes and part of the PAEs degradation gene cluster.

Based on these results, a possible conceptual model is proposed to describe the adaptive evolution of strain C1T (Figure 8B). Prior to exposure to e-waste contamination, the ancestor of strain C1T had at least a basic sequence skeleton of Chromosomes 1 and 2, likely harboring many oxygenase genes, the protocatechuate degradation gene cluster and a few heavy metal resistance genes. When initially exposed to e-waste contamination, i.e., an initial round of selection, it survived due to its natural resistance. The e-waste contaminated river sediment likely served as a “reservoir” of beneficial genes (Liu et al., 2018). Therefore, it had plenty of opportunities to acquire foreign plasmids and DNA fragments (mainly from sphingomonads) containing beneficial genes by transformation, conjugation and transduction. These plasmids and DNA fragments could exchange genes with the host genome by recombination, and/or be duplicated and inherited along the lineage. The high genome plasticity of strain C1T strongly suggested the possibility of these processes. As a result, intrinsic genes and acquired foreign genes could collectively form new composite metabolic characteristics. The accumulation of mobile elements including plasmids, prophages and ISs further increased genome plasticity. In addition, GD, genome rearrangements and sequence amelioration might occur. After subsequent rounds of selection, evolved strain C1T obtained a high genome plasticity and advantageous phenotype. The well-adapted strain C1T may now serve as a beneficial-gene contributor to promote the adaptive evolution of surrounding microorganisms and consequently accelerate the bioremediation process.

Future studies on experimental evolution of sphingomonads or communities and the relationship between genotype and phenotype will further illuminate the adaptive evolution mechanisms of sphingomonads for complex e-waste contamination.

DATA AVAILABILITY STATEMENT

The C1T genome sequences have been deposited in GenBank under the accession numbers CP022745, CP022746, CP022747, CP022748, CP022749, CP022750, and CP022751.

AUTHOR CONTRIBUTIONS

All authors conceived the study and revised the manuscript. DS, XC, MX, and RH designed the experiments. DS performed the experiments. DS, MX, AZ, and RT analyzed the data. DS wrote the manuscript.

FUNDING

This work was funded by the National Natural Science Foundation of China (91851202, S1678163, U1701243, and 21677042), GDAS’ Special Project of Science and Technology Development (2019GDASLY-0104005), Natural Science Foundation of Guangdong Province (2018B0303110010 and 2019B110205004), Science and Technology Project of Guangdong Province (2018B030324002), Science and Technology Project of Guangzhou (20170120021), and Guangdong technological innovation strategy of special funds (Key Areas of Research and Development Program) (2018B020205003).
ACKNOWLEDGMENTS

We thank Zhili He, Yinghua Cen, Jingjing Xu, Enze Li, and Rui Pang for advice on data analyses and manuscript revision.

REFERENCES

Alikhan, N., Petty, N. K., Ben Zakour, N. L., and Beatson, S. A. (2011). BLAST ring image generator (BRIG): simple prokaryote genome comparisons. BMC Genomics 12:402. doi: 10.1186/1471-2164-12-402

Andersson, D. I., and Hughes, D. (2009). Gene amplification and adaptive evolution in bacteria. Annu. Rev. Genet. 43, 167–195. doi: 10.1146/annurev-genet-102108-134805

Arndt, D., Grant, J. R., Marcu, A., Sajed, T., Pon, A., Liang, Y., et al. (2016). PHASTER: a better, faster version of the PHAST phage search tool. Nucleic Acids Res. 44, W16–W21. doi: 10.1093/nar/gkw387

Ayland, F. O., Mcdonald, B. R., Adams, S. M., Valenzuela, A., Schmidt, R. A., Goodwin, L. A., et al. (2013). Comparison of 26 sphingomonad genomes reveals diverse environmental adaptations and biodegradative capabilities. Appl. Environ. Microbiol. 79, 3724–3733. doi: 10.1128/AEM.00518-13

Barakat, M., Ortel, P., and Whitworth, D. E. (2013). P2RP: a web-based framework for the identification and analysis of regulatory proteins in prokaryotic genomes. BMC Genomics 14:269. doi: 10.1186/1471-2164-14-269

Bertelli, C. E. A. (2017). IslandViewer 4: expanded prediction of genomic islands for larger-scale datasets. Nucleic Acids Res. 45, W30–W35. doi: 10.1093/ngr/gks343

Chen, X., Song, D., Xu, J., Li, L., Sun, G., and Xu, M. (2017). Role and mechanism of cell-surface hydrophobicity in the adaptation of Sphingobium hydrophobicum to electronic-waste contaminated sediment. Appl. Microbiol. Biotechnol. 102, 2803–2815. doi: 10.1007/s00253-017-8734-2

Chen, X., Wang, H., Xu, J., Song, D., Sun, G., and Xu, M. (2016). Sphingobium hydrophobicum sp. nov., a hydrophobic bacterium isolated from electronic-waste contaminated sediment. Int. J. Syst. Evol. Microbiol. 66, 3912–3916. doi: 10.1099/ijsem.0.011287

Chin, C., Alexander, D. H., Marks, P., Klammer, A. A., Drake, J., Heiner, C., Fang, Y., Xu, M., Wu, W., Chen, X., Sun, G., Guo, J., et al. (2015). Characterization of Proteus mirabilis cell wall features in biofilm formation. Arch. Microbiol. 198, 877–884. doi: 10.1007/s00203-016-1249-x

Guo, P., and Dwan, B. (2017). Bacterial exopolysaccharide mediated heavy metal removal: a review on biosynthesis, mechanism and remediation strategies. Biotechnol. Rep. 13, 58–71. doi: 10.1016/j.btre.2016.12.006

Hemme, C. L., Green, S. J., Rishishwar, L., Prakash, O., Pettenato, A., Chakraborty, R., et al. (2016). Lateral gene transfer in a heavy metal-contaminated groundwater microbial community. mBio 7, e2234–215. doi: 10.1128/mBio.02234-15

Hu, J., Qian, M., Zhang, Q., Cui, J., Yu, C., Su, X., et al. (2015). Sphingobium fuliginis HC3: a novel and robust isolated biphenyl- and polychlorinated biphenyls-degrading bacterium without dead-end intermediates accumulation. PLoS One 10:e0122740. doi: 10.1371/journal.pone.0122740

Huertecapaz, J., Forsslund, K., Pedro, C. L., Szklarczyk, D., Juhl, J. L., Von, M. C., et al. (2017). Fast genome-wide functional annotation through orthology assignment by eggNOG-mapper. Mol. Biol. Evol. 34, 2115–2122. doi: 10.1093/molbev/msx148

Hyatt, D., Chen, G. L., Locascio, P. F., Land, M. L., Larimer, F. W., and Hauser, J. L. (2010). Prodigal: prokaryotic gene recognition and translation initiation site identification. BMC Bioinformatics 11:119. doi: 10.1186/1471-2105-11-119

Kanehisa, M., Sato, Y., and Morishima, K. (2015). BlastKOALA and GhostKOALA: KEGG tools for functional characterization of genome and metagenome sequences. J. Mol. Biol. 428, 726–731. doi: 10.1016/j.jmb.2015.11.006

Koonin, E. V., Makarova, K. S., and Aravind, L. (2001). Horizontal gene transfer in bacteria. Mol. Biol. Evol. 18:402. doi: 10.1186/1471-2164-18-402

Lagesen, K., Hallin, P., Redland, E. A., Staerfeldt, H. H., Rognes, T., and Ussery, D. W. (2007). RNAmmer: consistent and rapid annotation of ribosomal RNA genes. Nucleic Acids Res. 35, 3100–3108. doi: 10.1093/nar/gkm160

Laslett, D., and Canback, B. (2004). ARAGORN, a program to detect tRNA genes and tmRNA genes in nucleotide sequences. Nucleic Acids Res. 32, 11–16. doi: 10.1093/nar/gkh152

Leung, A., Cai, Z. W., and Wong, M. H. (2006). Environmental contamination in soil and plants on agricultural land near an electronic waste recycling site at Guiyu, southeast China. J. Mater. Cycles Waste Manage. 8, 1–14. doi: 10.1109/tpose.2008.925245

Lewis, M. R., and Logan, B. E. (2014). Microbial community structure and function in sediments from e-waste contaminated rivers at Guiyu area of China. Environ. Pollut. 193, 171–179. doi: 10.1016/j.envpol.2017.12.008

Liu, J., He, X., Lin, X., Chen, W., Zhou, Q., Xie, N., and He, X. (2015). Ecological effects of combined pollution associated with e-waste recycling on the composition and diversity of Soil microbial communities. Environ. Sci. Technol. 49, 6438–6447. doi: 10.1021/es5049804

Liu, X., Gai, Z., Tao, F., Tang, H., and Xu, P. (2012). Carotenoids play a positive role in the degradation of heterocycles by Sphingobium yanoikuyae. PLoS One 7:e39522. doi: 10.1371/journal.pone.0039522

Ma, T. T., Christie, P., Luo, Y. M., and Teng, Y. (2013). Phthalate esters in the microbial cell wall of Proteus mirabilis. Curr. Microbiol. 63, 1335–1337. doi: 10.1007/s00284-013-0939-5

Nevo, E. (2011). Evolution under environmental stress at macro- and microscales. Genome Biol. Evol. 3, 1039–1052. doi: 10.1093/gbe/evr052

Grzegorczyk, C., Anna, G., Klaudia, K., Michalina, G., Magdalena, D., Lukasz, L., et al. (2016). The role of Proteus mirabilis cell walls features in biofilm formation. Arch. Microbiol. 198, 877–884. doi: 10.1007/s00203-016-1249-x

REFERENCES SUPPLEMENTARY MATERIAL

The Supplementary Material for this article can be found online at: https://www.frontiersin.org/articles/10.3389/fmicb.2019.02263/full#supplementary-material
Tabata, M., Ohhata, S., Nikawadori, Y., Kishida, K., Sato, T., Kawasumi, T., Stolz, A. (2009). Molecular characteristics of xenobiotic-degrading Rodger, K. R., and Lukat-Rodgers, G. S. (2005). Insights into heme-based O2 Richter, M., Rosselló-Móra, R., Oliver Glöckner, F., and Peplies, J. (2016). Qu, Y., Wang, J., Zhou, J., and Xing, L. (2006). Decolorization of bromoamine acid Nörtemann, B., Baumgarten, J., Rast, H. G., and Knackmuss, H. J. (1986). Bacterial Song et al. Adaptive Evolution of Varani, A. M., Siguier, P., Gourbeyre, E., Charneau, V., and Chandler, M. (2011). ISsaga is an ensemble of web-based methods for high throughput identification and semi-automatic annotation of insertion sequences in prokaryotic genomes. Genome Biol. 12:R30. doi: 10.1186/gb-2011-12-3-30 Verha, H., Kumar, R., Oldach, P., Sangwan, N., Khurana, J. P., Gilbert, J. A., et al. (2014). Comparative genomic analysis of nine Sphingobium strains: insights into their evolution and hexachlorocyclohexane (HCH) degradation pathways. BMC Genomics 15:1014. doi: 10.1186/1471-2164-15-1014 Vesth, T., Lagesen, K., Acar, Ö., and Usery, D. (2013). CMG-Biotools, a free workbench for basic comparative microbial genomics. PLoS One 8:e60120. doi: 10.1371/journal.pone.0060120 Wang, D., Zhang, Y., Zhang, Z., Jiang, Z., and Yu, J. (2010). KaKs_Calculator 2.0: a toolkit incorporating gamma-series methods and sliding window strategies. Genomics Proteomics Bioinformatics 8, 77–80. doi: 10.1016/S1672-0229(10)60008-3 Wang, J., Wang, C., Li, J., Bai, P., Li, Q., Shen, M., et al. (2018). Comparative genomics of degradative Novosphingobium strains with special reference to microcytstin-degrading Novosphingobium sp. THN1. Front. Microbial. 9:2238. doi: 10.3389/fmicb.2018.02238 Wang, Y., Tang, H., Debarry, J. D., Tan, X., Li, J., Wang, X., et al. (2012). MCScanX: a toolkit for detection and evolutionary analysis of gene synteny and collinearity. Nucleic Acids Res. 40:e49. doi: 10.1093/nar/gkr1293 Yang, H. C., and Rosen, B. P. (2016). New mechanisms of bacterial arsenic resistance. Biomed. J. 39, 5–13. doi: 10.1016/j.bj.2015.08.003 Yee, N., Fein, J. B., and Daughney, C. J. (2000). Experimental study of the pH, ionic strength, and reversibility behavior of bacteria-mineral adsorption. Geochim. Cosmochim. Acta 64, 609–617. doi: 10.1016/S0016-7037(99)00342-7 Zhang, K., Schnoor, J. L., and Zeng, E. Y. (2012). E-waste recycling: where does it go from here? Environ. Sci. Technol. 46, 10861–10867. doi: 10.1021/es303166s Zhang, Z., Xiao, J., Wu, J., Zhang, H., Liu, G., Wang, X., et al. (2012). ParaAT: a parallel tool for constructing multiple protein-coding DNA alignments. Biochem. Biophys. Res. Commun. 419, 779–781. doi: 10.1016/j.bbrc.2012.02.101 Conflict of Interest: The authors declare that the research was conducted in the absence of any commercial or financial relationships that could be construed as a potential conflict of interest. Copyright © 2019 Song, Chen, Xu, Hai, Zhou, Tian, Van Nostrand, Kempher, Guo, Sun and Zhou. This is an open-access article distributed under the terms of the Creative Commons Attribution License (CC BY). The use, distribution or reproduction in other forums is permitted, provided the original author(s) and the copyright owner(s) are credited and that the original publication in this journal is cited, in accordance with accepted academic practice. No use, distribution or reproduction is permitted which does not comply with these terms.