Pruning from spheroidal to cubic canopy induced aphid outbreak by altering the plant performance in spindle tree

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DOI:  
10.21203/rs.2.12158/v1

SUBJECT AREAS  
Terrestrial Ecology

KEYWORDS  
aphid, herbivore, plant-animal interactions
Abstract

The effects of pruning on animal-plant interactions have rarely been studied. We thus conducted field experiments to examine the impact of artificially-pruned shapes (e.g. cubic and spheroidal canopy) on the performance of the spindle tree and the resulting aphid abundance at three sites; on a university campus, at a road green belt, and in a residential area. The results showed aphid abundance was 89.9 times higher in plants with a cubic canopy than with a spheroidal canopy. Plants with a cubic canopy had lower leaf dry mass content and inflorescence numbers, but greater fresh twig length than the plants with a spheroidal canopy. The aphid abundance was negatively correlated with the leaf dry mass content and inflorescence numbers, and positively correlated with the fresh twig length. The results can provide data support for human actives can alter plant performance, and thereby to change insect preference.

Introduction

Plant-animal interactions comprise the fundamental relationships of ecological research (Blue et al. 2011, Li et al. 2011, Liu et al. 2011, Martin and Maron 2012, Fang and Huang 2014, Mu et al. 2014). Changes in the relationships between animals and plants can affect plant growth and reproduction (Liu et al. 2011, Fang and Huang 2014, Mu et al. 2014, Shao et al. 2017), aboveground biomass (Blue et al. 2011, Li et al. 2011), community composition (Strauss and Irwin 2004, Nogales et al. 2015), nutrient cycling and energy flux (Heimann and Reichstein 2008, Defossez et al. 2011). Most previous studies have concentrated on effects of environmental change on plant-animal interactions (Morris et al. 2008, Malhi et al. 2009, Li et al. 2011, Liu et al. 2011, de Sassi et al. 2012, Hoover et al. 2012, Martin and Maron 2012) and have found that these relationships could be affected by rising temperatures (Li et al. 2011, Liu et al. 2011), changes in precipitation
regimes (Morris et al. 2008, Malhi et al. 2009) and differential nitrogen deposition (de Sassi et al. 2012, Hoover et al. 2012). However, the effects of human activity also affect plant-animal interactions (Romero et al. 2006, Rosin and Poulsen 2016), especially in the case of plant and pollinator mutualisms (Aizen and Feinsinger 1994, Kearns et al. 1998). Pruning is one widespread human activity that directly affects a plant’s ability to perform vital life processes, such as photosynthesis (Agrawal and Spiller 2004, Grechi et al. 2008). Plant traits, including nutrient content and physical or chemical defenses, may also be altered by pruning (Martinez and Wool 2002, Agrawal and Spiller 2004, Kumar et al. 2010, Saifuddin et al. 2010, Maltoni et al. 2012, Albarracín et al. 2017) and could subsequently affect a plant’s accessibility or attractiveness to herbivores (Nykänen and Koricheva 2004, Wen et al. 2006). However, the effects of pruning on the appetite of herbivores are inconsistent in previous studies, and thus unknown. On one hand, light pruning may remove unhealthy tissue, and reduce pest damage. This is known as the “escape strategy” (Rattan 1992, Sivapalan 1999, Maltoni et al. 2012). On the other hand, pruning produces vigorous shoots and leaves, which could attract more herbivores. This is called the plant vigor hypothesis (Price 1991, Grechi et al. 2008). Furthermore, a plant undergoing a heavy pruning treatment produces more tender twigs and leaves, which leads to a higher water content in the plant than under undisturbed conditions. The tender twigs could attract fewer pests, according to the plant-stress hypothesis (White 1969), or it could attract more sap-feeders, according to the pulsed stress hypothesis (Huberty and Denno 2004). Few studies have tested these theories and investigated the difference in predator abundance between lightly- and heavily-pruned plants of the same species, or the influence of pruning on plant structure (e.g., leaf mass, twig length etc.) within a species. We will use these metrics to assess the impact of pruning on herbivorous insect abundance and to increase our understanding of the response mechanisms of trophic
relationships to human activity. Considering both that heavy pruning induces more compensatory growth than does light pruning and that plant structure and insect attraction are tightly correlated, we hypothesized that heavy pruning would produce more vigorous shoot and leaf growth and thereby attract more sap-feeders.

Our study system consists of the aphid species (*Aphis gossypii* Glover) and one of its host species, the spindle tree (*Buxus megistophylla* Levl.). The spindle tree is a popular evergreen garden plant and is cultivated in most provinces of China. The spindle tree is often pruned to one of two canopy shapes: a cubic canopy, like a hedge with heavy pruning and a spheroidal canopy, like a ball with light pruning. We investigated the abundance of aphids, and we measured plant structure with the leaf dry mass content, the length of the annual shoots and the inflorescence number of the differently-shaped plants. Our objectives were 1) to test whether the different shapes of the plants would affect the plant-animal interactions, and 2) to explore the response of animals to different branch pruning shapes in plants they often eat.

**Materials And Methods**

**Study site**

In 2015, three adjacent 1 km × 1 km sites were selected for this study, a university campus (Jinming campus district of Henan University, 34°49′ 17″N, 114° 17′ 57″E, 73 m), a road green belt (Dongjing Road, 34°49′25″N, 114°18′19″E, 73 m) and a residential area (Longxiang-Shangrila district, 34°49′33″N, 114°18′19″E, 73 m) in the northwest district of Kaifeng city in Henan Province, China. In the study area, the mean annual precipitation is 626 mm and the mean annual temperature is 14 ℃ (Zhang et al. 2009). The soil is sandy loam, according to the FAO classification system (FAO 1990). The vegetation on the three sites is of a manmade composition, with the spindle tree (*Buxus megistophylla* Levl.) as
the dominant species. The plots were selected followed with located observation method (Miao et al. 2018). In each study site, 10 pairs of plots were randomly selected, where each pair contained \emph{B. megistophylla} Levl. having both cubic canopy (CU) and spheroidal canopy (SP). A consistent frequency and timing of disturbances, which included pruning, irrigation, and application of pesticide, were maintained for each paired-plot within a single site. Each pair of plots was set apart from other pairs at a distance of greater than 10 m.

**Study species**

\emph{Buxus megistophylla} Levl. was chosen as the plant study subject. \emph{B. megistophylla} Levl. is a common evergreen garden shrub in China, often subjected to frequent pruning to maintain its manicured shape. The shrubs in this study were planted in years 2006, 2010, and 2013 at the university campus, the road green belt and the residential area, respectively. The shrubs often produce a large number of fresh twigs in the spring and were pruned irregularly from late spring to late summer at the study sites. Shrubs were pruned to have either a cubic canopy (CU) or a spheroidal canopy (SP) at all three study sites (Fig. 1). Additionally, the shrubs were watered semi-monthly and sprayed irregularly with omethoate pesticide from late May to late September of each growing year. The distance between adjacent CU and SP shrubs was 3 m, 1 m, and 5 m at the university campus, road green belt and residential area sites, respectively. Large numbers of aphids were found on the shrubs with a CU in the year 2015, thus we selected this species. Aphids are the most common pest of \emph{B. megistophylla} Levl. They first appeared on the shrub each year in late April at our study sites. After capturing a sample of aphids, fixing them with 75% alcohol and observing them under a stereoscopic microscope with an eyepiece micrometer, we identified them as cotton aphids (\emph{Aphis gossypii} Glover),
according to the telltale length of their cornicles relative to their caudas, the shapes of
their caudas, and their body color, as described in the Fauna of Hebei, China, Aphidinea
(Qiao et al. 2009). The cotton aphid is a holocyclic species (Qiao et al. 2009). They often
lay their eggs near the leaf bud of their overwintering host, which may be a grape vine
(*Vitis vinifera* L.), pomegranate plant (*Punica granatum* L.) or spindle tree (*B.
megistophylla* Levl.), and they remain dormant during the cold winter until they hatch in
early spring. They reproduce several generations and then migrate to cotton plants to
breed again. Cotton aphids suck the juice of fresh leaves and twigs, often inhibiting plant
growth.

**Aphid abundance measurements**
The investigation of aphid abundance was conducted on one day, April 26, 2015. Three
annual twigs representing both CU and SP were randomly collected from one individual,
and three individuals of CU and SP were randomly collected in each plot. The abundance
of aphids on each of the selected twigs was counted with the naked eye. The aphid
abundance on the selected twigs was first averaged for both CU and SP plants on each
individual, and then averaged in each plots. The herbivore load was calculated to
determine whether the difference in abundances could be attributed to either the
difference in plant structure or to an attraction to one of the two shrub shapes. Herbivore
load was calculated as the ratio between the number of the aphids and the length of the
twig.

**Plant structural measurements**
The twigs selected for the aphid abundance measurements were clipped from the shrubs
and measured after the aphids were counted. The leaves on each twig were collected and
weighed first at their fresh weight, and then again at their dry weight after being oven-
dried to a constant weight at 65 °C. Leaf dry mass content (DMC) was calculated as the ratio between the dry weight and the fresh weight as a metric for the tenderness of the leaf, which is closely correlated with attractiveness to aphids.

A month after the aphid abundance measurements, three other perennial twigs were randomly selected in the selected individuals. The number of inflorescences in each selected perennial twig was recorded.

**Statistical analysis**

Normal distributions were tested for each parameter before the data analysis. Aphid abundance and twig length were log-transformed to meet the assumptions of ANOVA. Two-way ANOVAs were conducted to compare plant structure and aphid abundance at the different sites and with the different pruning shapes. A paired-t test was employed to determine the effect of pruning shape on plant performance and aphid abundance and simple linear regressions were run to determine their relationship. All the above statistics were calculated using SPSS 19.0 software package (SPSS Inc., CHI, IL, USA).

In addition, a structural equation model (SEM) was employed to detect causal linkages from explanatory variables to aphid abundance, including plant structural parameters following a previous study (Sun et al. 2016). The strength of direct and indirect relationships among all the considered variables was estimated with linear regression results (Grace 2006). Pairs of variables were checked for bivariate relationships before establishing the linear models to eliminate confounding factors. The SEM models were constructed based on the known effects and potential relationships among the aphid abundance and the plant modular parameters. Data was fitted to the models using the maximum likelihood estimation method. The $\chi^2$-test was used to evaluate the fitness of each model (Grace and Bollen 2005). All SEM analyses were performed with AMOS 18.0
Results

**Aphid abundance**

Aphid abundance varied greatly with study site. The aphid abundance at the road green belt was significantly higher than at the campus or the residential area sites (ANOVA: F = 36.3, Table 1; Fig. 2a). Pruning shape was found to significantly affect aphid abundance at all three sites (ANOVA: F = 358.8, Table 1). At each site, more than 30 aphids were found on each stem from a CU, whereas fewer than 5 aphids were observed on each stem from an SP. Aphid abundance on the CU was 89.9 times higher than it on the SP across all the three sites. Specifically, aphid abundance on the CU was 46.7, 113.5 and 37.5 times higher than on the SP at the campus, road green belt and residential area sites, respectively (t-test: t = 5.1, t = 7.9, t = 3.5, Table 2).

The herbivore load on the CU was 22.9 times larger than on the SP across the three sites (t-test: t = 5.7, t = 8.2, t = 3.7, Table 2; Fig.2b). Additionally, the effect of pruning shape on the herbivore load varied significantly with the study site (ANOVA: F = 35.1, Table 1). The herbivore load of the CU was 32.2 times larger than of the SP at the road green belt site, but only 12.9 and 12.8 times larger than of the SP at the campus and residential area sites, respectively.

**Plant modular Parameters**

Twig length varied significantly with study site (ANOVA: F = 547.1, Table 1), with the greatest (9.60 cm) and shortest (3.85 cm) lengths occurring at the road green belt and residential area sites, respectively (t-test: t = 12.4, t = 8.4, Table 2; Fig. 3a). The twig length in the CU was on average 6.2 cm longer than in the SP across the three sites (Fig. 3a). Further, the effect of the pruning shape on twig length varied with study site (ANOVA:
The twig length of the CU samples were 4.4, 10.6, and 3.7 cm longer than the SP samples at the campus, road green belt and residential area sites, respectively.

The leaf dry mass content (LDWC) was significantly affected by the study site (ANOVA: $F = 37.4$, Table 1). The LDWC at the residential area site was lower than at the other two sites (Fig. 3b). The LDWC of the CU samples (27.0%) was significantly lower than for the SP samples (34.3%) across the three sites. The site and pruning shape interacted to affect LDMC in the experiment (ANOVA: $F = 11.1$, Table 1). The LDWC of the CU samples was 8.54%, 5.01%, and 8.55% lower in absolute change than of the SP samples from the campus, road green belt and residential area sites, respectively (t-test: $t = 15.2$, $t = 6.3$, $t = 20.8$, Table 2).

The inflorescence number was significantly affected by study site (ANOVA: $F = 23.5$, Table 1), with the maximum and minimum values occurring at the campus and the road green belt sites, respectively (Fig. 3c). The inflorescence number varied significantly with the pruning shape (ANOVA: $f = 775.0$, Table 1). The inflorescence number on the CU was 16.3 times lower than on the SP (Fig. 3c). The effect of pruning shape on inflorescence number varied with study site (ANOVA: $f = 17.3$, Table 1). The inflorescences on the SP numbered 16.5, 20.3 and 14.1 times greater than on the CU at the campus, road green belt and residential area sites, respectively (t-test: $t = 20.4$, $t = 15.1$, $t = 15.0$, Table 2).

**Relationships between aphid abundance and plant structural parameters**

The aphid number decreased with increasing LDMC at all study sites (all $P < 0.001$; Fig. 4a-c). However, when the regression analysis was conducted for each the CU and the SP
separately, aphid abundance on CU was the only factor negatively correlated with LDMC, and only at one site, the road green belt (Fig. 4e); no other relationships between aphid abundance and leaf DMC were detected at either study site (all \( P > 0.05 \); Fig. 4d; 4f-i). The aphid abundance increased with increasing twig length linearly at all study sites (all \( P < 0.001 \); Fig. 5a-c). When the regression analysis was conducted for each the CU and the SP separately, the aphid abundance in the CU was positively correlated with twig length at all the sites (all \( P < 0.01 \); Fig. 5d-f). However, the aphid abundance in the SP was not correlated with twig length at any of the three sites (all \( P > 0.05 \); Fig. 5g-i).

Aphid abundance decreased linearly with increasing inflorescence number at all study sites (all \( P < 0.001 \); Fig. 6a-c). When the regression analysis was conducted for each the CU and the SP separately, the aphid abundance in the CU was negatively correlated with inflorescence number at the campus and the road green belt sites (all \( P < 0.05 \); Fig. 6d-e), but not with inflorescence number at the residential area site (Fig. 6f). In addition, the aphid abundance in the SP was negatively correlated with inflorescence number at all three sites (all \( P < 0.05 \); Fig. 6g-i).

The best SEM model (\( \chi^2 = 2.797, P = 0.094, \text{df} = 1 \)) explained 91% of the variation in aphid abundance. The variation can primarily be attributed to the increase in fresh twig length (Fig. 7). Light pruning (from CU to SP) and the resulting negative effect on twig length led to a decrease in aphid abundance. Light pruning also positively and significantly affected leaf dry mass content and inflorescence number, although the SEM indicated that the increase in leaf dry mass content or inflorescence number barely accounted for the variation in aphid abundance.

**Discussion**

Our results show that the plants with a cubic canopy had higher aphid abundance, lower
leaf dry mass content and lower inflorescence number, but they also had greater fresh
twig length than plants with a spheroidal canopy. These results support our hypothesis
that heavy pruning can significantly augment an herbivorous population through the
alteration of plant performance, and suggest that the pruning-induced changes of plant
performance may result in the variations in herbivores.

Pruning is an important horticultural management method, and has been reported to alter
canopy size, to elevate growth rate and to increase the number of shoots to compensate
for structures lost in pruning (Albert et al. 2010). Different pruning strategies have
different effects on plant morphology (Tworkoski et al. 2006, Mampionona 2011). For
instance, Mampionona (2011) reported that trees with a long trunk and short primary
branches that were pruned had longer lateral shoots compared to the post-pruning growth
from a tree with a short trunk and long primary branches. Tworkoski, Miller, and Scorza
(2006) reported that pillar trees exhibit more upright growth after pruning, but have fewer
sylleptic branches than standard trees. In the current study, we have observed that plants
with a spheroidal canopy allocate more resources to reproductive (e.g., inflorescence),
rather than vegetative, growth (e.g., twig length) after pruning than plants with a cubic
canopy. These different performances could be attributed to the heavier pruning
treatment necessary for plants to maintain a cubic canopy rather than a spheroidal
canopy. The former can produce many epitomic shoots, according to the compensating
theory (Guilett and Bergström 2006). After the apical tissue of the shoots was pruned, the
plants with a cubic canopy maintained dense shoots at the surface. High-density plants,
like these with a cubic canopy, must grow sufficiently upward to obtain enough light (Vile
et al. 2006, Liu et al. 2012). Because the spheroidal canopy is similar in shape to the
natural form of shrubs, plants with a spheroidal canopy are also not subject to light
limitation for growth in the same way as shrubs with a cubic canopy. Most resources of
shrub with a spheroidal canopy are spent on new outward growth, rather than the re-
growth of twigs. As a result, plants undergoing the less intensive pruning regimen for
spheroidal canopies are much weaker than plants with a cubic canopy. Our results are
consistent with previous findings for litchi (Olesen et al. 2013) and blueberry (Lee et al.
2015), which indicate that different pruning strategies may redistribute the trade-off
between vegetative and reproductive growth in many species.

This study has found that moderate pruning toward a spheroidal canopy shape results in a
lower abundance of insect herbivores than heavy pruning toward a cubic canopy shape,
which supports our hypothesis. Our results are consistent with previous studies that have
reported that pruning may attract more sap-feeders (Price 1991, Grechi et al. 2008).
However, our results contrast the results of previous studies (Cornelius 2001, Simon et al.
2007, Mdellel et al. 2015) that have reported that pruning may reduce the herbivory
behavior of insects (Cornelius 2001, Simon et al. 2007, Mdellel et al. 2015) due to the
removal of terminal shoots (Martinez and Wool 2003) and the reduction of the insect eggs
during pruning. The lower abundance of aphids on the spheroidal canopy shrubs can be
largely attributed to the shorter, regenerated branches of the plants, when compared to
shrubs with a cubic canopy. As a common sap-feeder, the aphid often lives on tender
buds, leaves, and twigs (Chau and Mackauer 1997, Hu et al. 2017). No aphids were found
on our plants with either canopy form immediately after either pruning treatment, thus
any aphids found on the shrubs after pruning had migrated from other plants (Observed
by Chen). Our results for herbivore load indicate that the twigs on stems in cubic canopies
attracted more aphids than the stems in spheroidal canopies because the plants with
cubic canopies had longer, fresher twigs and lower leaf dry mass content. Alternatively,
the plant with the spheroidal canopy produced more sap, which may have attracted more
aphids, as stated in the pulsed stress hypothesis (Huberty and Denno 2004). Shrubs with
the cubic canopy may also have been preferential for the herbivores because the longer, fresher twig can also access more sunlight and thus maintain a higher, more comfortable temperature. Barton and Ives (2014) reported that aphid abundance should increase at higher temperature where predators are absent. No predator were found in our study, thus we can expected the longer, fresher twigs on the plants with a cubic canopy to attract a larger abundance of aphids than the plants with a spheroidal canopy. To test this hypothesis, we observed aphid abundance on plants with both a cubic and a spheroidal canopy in the road green belt both prior to and following the first pruning treatment. The aphids disappeared after the first pruning. This finding suggests that the initial observations of aphids are most accurate and that the aphids had a higher reproductive rate among the cubic canopies (Appendix S1: Fig. S1). Furthermore, previous studies have reported that plants with different structures contain different concentrations of secondary metabolites, such as flavonoids (Conde et al. 2009, Maudu et al. 2010), which could affect the survival of insects that consume plant parts (Lattanzio et al. 2000). However, we did not detect significant differences in the flavonoid content between the plants having cubic and spheroidal canopies (Appendix S2: Fig. S2), thus the influence of the flavonoids on the aphids was considered negligible in the study.

Our results have demonstrated that the influence of pruning on aphid abundance significantly varies with study site, which indicates that significant interactions are at play between pruning and study site on aphid abundance. Two possible reasons may be employed to explain the interactions. First, the road green belt is more open than the other two sites, thus the growth rate of the aphid population at the road green belt may be greater than at the other two sites. Actually, the aphid abundance in the cubic canopy stems at the road green belt was also much greater than at the other two sites, but the difference in aphid abundance in the spheroidal canopy stems among the three sites was
smaller than for the abundance in the cubic canopy only. Secondly, the horticulture management (e.g., irrigation, pruning, and pesticide application) differed in frequency among the three sites. According to our observations, the horticulture management was less frequent at the road green belt than at the other two sites, which may have induced the larger population of aphids. Our results indicate that the pruning effects may be altered according to the environment and other horticulture management practices.

Conclusion

Our results suggest that (1) aphid abundance on plants with a cubic canopy is higher than on the plants with a spheroidal canopy and that (2) aphid abundance is mainly correlated with fresh twig length and inflorescence number. Our findings have proven that pruning shape can significantly affect the abundance of herbivores on the pruned plants. The results can provide data to support for the human activities can alter plant modular that affect insect preference.

Declarations

Ethics approval and consent to participate

Not applicable.

Consent for publication

We understand that the information will be published without our/our child or ward’s/my relative’s (circle as appropriate) name attached, but that full anonymity cannot be guaranteed.

We understand that the text and any pictures or videos published in the article will be freely available on the internet and may be seen by the general public. The pictures, videos and text may also appear on other websites or in print, may be translated into other languages or used for commercial purposes.

We have been offered the opportunity to read the manuscript there is no financial interest
Availability of data and material
All data used in this study are included in the manuscript and supporting information.

Competing interests
The authors declare no conflict of interest

Funding
This study was supported by the National Natural Science Foundation of China (31670477, 31200375), the China Postdoctoral Science Foundation (2017M612389) and the outstanding youth training foundation of Henan University (yqpy20140031).

Author contributions
Yinzhan Liu designed the experiment and wrote the manuscript. Juan Xuan and Anqun Chen collected the data, and conducted the data analysis. Chunlian Qiao and Renhui Miao revised the manuscript. Yinzhan Liu and Juan Xuan contributed equally in the work.

Acknowledgements
The authors wish to thank Xinqiang Xi and Yanchun Liu for their valuable suggestions regarding experimental design and manuscript revision. Thanks for Elizabeth Tokarz at the Yale University, who worked in the Mapleedit Company in Tianjin for the English revision.

Reference
Agrawal, A. A., and D. A. Spiller. 2004. Polymorphic buttonwood: effects of disturbance on resistance to herbivores in green and silver morphs of a Bahamian shrub. American Journal of Botany 91:1990-1997.

Aizen, M. A., and P. Feinsinger. 1994. Forest fragmentation, pollination, and plant reproduction in a Chaco dry forest, Argentina. Ecology 75:330-351.

Albarracín, V., Hall, A. J., P. S. Searles, and M. C. Rousseaux. 2017. Responses of
vegetative growth and fruit yield to winter and summer mechanical pruning in olive trees. Scientia Horticulturae 225:185-194.

Albert, T., Karp, K., M. Starast, and T. Paal. 2010. The effect of mulching and pruning on the vegetative growth and yield of the half-high blueberry. Agronomy research 8:759–769.

Barton, B. T., and A. R. Ives. 2014. Direct and indirect effects of warming on aphids, their predators, and ant mutualists. Ecology 95:1479-1484.

Blue, J. D., L. Souza, A. T. Classen, J. A. Schweitzer, and N. J. Sanders. 2011. The variable effects of soil nitrogen availability and insect herbivory on aboveground and belowground plant biomass in an old-field ecosystem. Oecologia 167:771-780.

Chau, A., and M. Mackauer. 1997. Dropping of pea aphids from feeding site: a consequence of parasitism by the wasp, Monoctonus paulensis. Entomologia experimentalis et applicata 83:247-252.

Conde, E., C. Cara, A. Moure, E. Ruiz, E. Castro, and H. Domínguez. 2009. Antioxidant activity of the phenolic compounds released by hydrothermal treatments of olive tree pruning. Food Chemistry 114:806-812.

Cornelius, J. P. 2001. The effectiveness of pruning in mitigating Hypsipyla grandella attack on young mahogany (Swietenia macrophylla King) trees. Forest Ecology and Management 148:287-289.

de Sassi, C., O. T. Lewis, and J. M. Tylianakis. 2012. Plant-mediated and nonadditive effects of two global change drivers on an insect herbivore community. Ecology 93:1892-1901.

Defossez, E., C. Djieto-Lordon, D. McKey, M. A. Selosse, and R. Blatrix. 2011. Plant-ants feed their host plant, but above all a fungal symbiont to recycle nitrogen. Proceedings of the Royal Society of London B: Biological Sciences 278:1419-1426.

Fang, Q., and S. Huang. 2014. Progress in pollination ecology at the community level.
Chinese Science Bulletin 59:449–458.

FAO, 1990. Guidelines for Soil Descriptions. 3rd ed. FAO, Rome.

Grace, J. B. 2006 Structural equation modeling and natural systems. Cambridge University Press, Cambridge.

Grace, J. B., and K. A. Bollen. 2005 Interpreting the results from multiple regression and structural equation models. Bulletin of the Ecological Society of America 86:283–295.

Grechi, I., M. H. Sauge, B. Sauphanor, N. Hilgert, R. Senoussi, and F. Lescourret. 2008. How does winter pruning affect peach tree-Myzus persicae interactions? Entomologia Experimentalis et Applicata 128:369–379.

Guillet, C., and R. Bergström. 2006. Compensatory growth of fast-growing willow (Salix) coppice in response to simulated large herbivore browsing. Oikos 113:33–42.

Heimann, M., and M. Reichstein. 2008. Terrestrial ecosystem carbon dynamics and climate feedbacks. Nature 451:289–292.

Hoover, S. E. R., J. J. Ladley, A. A. Shchepetkina, M. Tisch, S. P. Gieseg, and J. M. Tylianakis. 2012. Warming, CO2, and nitrogen deposition interactively affect a plant-pollinator mutualism. Ecology Letters 15:227–234.

Hu, D. W., S. Zhang, J. Y. Luo, L. M. Lu, J. J. Cui, and X. Zhang. 2017. An example of host plant expansion of host-specialized Aphis gossypii Glover in the field. Plos One 12:13.

Huberty, A. F., and R. F. Denno. 2004. Plant water stress and its consequences for herbivorous insects: a new synthesis. Ecology 85:1383–1398.

Kearns, C. A., D. W. Inouye, and N. M. Waser. 1998. Endangered mutualisms: the conservation of plant-pollinator interactions. Annual review of ecology and systematics 29(1):83–112.

Kumar, M., V. Rawat, J. M. S. Rawat, and Y. K. Tomar. 2010. Effect of pruning intensity on peach yield and fruit quality. Scientia horticulturae 125:218–221.
Lattanzio, V., S. Arpaia, A. Cardinali, D. Di Venere, and V. Linsalata. 2000. Role of endogenous flavonoids in resistance mechanism of Vigna to aphids. Journal of agricultural and food chemistry 48:5316-5320.

Lee, S. G., J. G. Cho, M. H. Shin, S. B. Oh, H. L. Kim, and J. G. Kim. 2015. Effects of summer pruning combined with winter pruning on bush growth, yields, and fruit quality of ‘Misty’ southern highbush blueberry for two years after planting. Horticulture, Environment, and Biotechnology 56:740-748.

Li, G., Y. Liu, L. E. Frelich, and S. Sun. 2011. Experimental warming induces degradation of a Tibetan alpine meadow through trophic interactions. Journal of Applied Ecology 48:659-667.

Liu, Y., J. Mu, K. J. Niklas, G. Li, and S. Sun. 2012. Global warming reduces plant reproductive output for temperate multi-inflorescence species on the Tibetan plateau. New Phytologist 195:427-436.

Liu, Y., P. B. Reich, G. Li, and S. Sun. 2011. Shifting phenology and abundance under experimental warming alters trophic relationships and plant reproductive capacity. Ecology 92:1201-1207.

Malhi, Y., L. E. Aragao, D. Galbraith, C. Huntingford, R. Fisher, P. Zelazowski, S. Sitch, C. McSweeney, and P. Meir. 2009. Exploring the likelihood and mechanism of a climate-change-induced dieback of the Amazon rainforest. Proceedings of the National Academy of Sciences 106:20610-20615.

Maltoni, A., B. Mariotti, D. F. Jacobs, and A. Tani. 2012. Pruning methods to restore Castanea sativa stands attacked by Dryocosmus kuriphilus. New forests 43:869-885.

Mampionona, R. A. 2011. Physiological and growth responses of Jatropha curcas L. to water, nitrogen and salt stresses. Dissertation, Institute of plant production and agroecology in the tropics and subtropics, university of Hohenheim, Stuttgart, Baden-
Württemberg, Germany

Martin, T. E., and J. L. Maron. 2012. Climate impacts on bird and plant communities from altered animal-plant interactions. Nature Climate Change 2:195–200.

Martinez, J. J. Y., and D. Wool. 2003. Differential response of trees and shrubs to browsing and pruning: the effects on Pistacia growth and gall-inducing aphids. Plant Ecology 169:285–294.

Maudu, M., F. N. Mudau, and I. K. Mariga. 2010. The effect of pruning on growth and chemical composition of cultivated bush tea (Athrixia phyllicoides DC). Journal of Medicinal Plants Research 4:2353–2358.

Mdellel, L., H. M. B. Kamel, and B. Assadi. 2015. Impact of winter pruning of pomegranate trees on Aphis punicae (Hemiptera, Aphididae) and its natural enemies in Tunisia. International Journal of Entomology 51:266–271.

Miao, R. H., X. L. Qiu, M. X. Guo, A. Musa, and D. M. Jiang. 2018. Accuracy of space-for-time substitution for vegetation state prediction following shrub restoration. Journal of Plant Ecology 11:208–217.

Morris, W. F., C. A. Pfister, S. Tuljapurkar, C. V. Haridas, C. L. Boggs, M. S. Boyce, E. M. Bruna, D. R. Church, T. Coulson, D. F. Doak, S. Forsyth, J. M. Gaillard, C. C. Horvitz, S. Kalisz, B. E. Kendall, T. M. Knight, C. T. Lee, and E. S. Menges. 2008. Longevity can buffer plant and animal populations against changing climatic variability. Ecology 89:19–25.

Mu, J., Y. Peng, X. Xi, X. Wu, J. N. Griffin, K. J. Niklas, and S. Sun. 2014. Domesticated honey bees evolutionarily reduce flower nectar volume in a Tibetan lotus. Ecology 95:3161–3172.

Nogales, M., I. Castañeda, M. López-Darias, F. M. Medina, and E. Bonnaud. 2015. The unnoticed effect of a top predator on complex mutualistic ecological interactions. Biological Invasions 17:1655–1665.
Nykänen, H., and J. Koricheva. 2004. Damage-induced changes in woody plants and their effects on insect herbivore performance: a meta-analysis. Oikos 104:247-268.

Olesen, T., C. M. Menzel, C. A. McConchie, and N. Wiltshire. 2013. Pruning to control tree size, flowering and production of litchi. Scientia Horticulturae 156:93-98.

Price, P. W., 1991. The plant vigor hypothesis and herbivore attack. Oikos 62:244-251.

Qiao, G., G. Zhang, L. Jiang, T. Zhong, and S. Tian. 2009 The fauna of Hebei, China, Aphidinea. Shijiazhuang: Hebei Science and Technology Publishing House (in Chinese).

Rattan PS. 1992. Pest and disease control in Africa. In Wilson KC, Clifford MN, eds. Tea: Cultivation to Consumption. London: Chapman and Hall.

Romero, G. Q., P. Mazzafera, J. Vasconcellos-Neto, and P. C. Trivelin. 2006. Bromeliad-living spiders improve host plant nutrition and growth. Ecology 87:803-808.

Rosin, C., and J. R. Poulsen. 2016. Hunting-induced defaunation drives increased seed predation and decreased seedling establishment of commercially important tree species in an Afrotropical forest. Forest Ecology and Management 382:206-213.

Saifuddin, M., A. B. M. S. Hossain, N. Osman, M. A. Sattar, K. M. Moneruzzaman, and M. I. Jahirul. 2010. Pruning impacts on shoot-root-growth, biochemical and physiological changes of ‘Bougainvillea glabra’. Australian Journal of Crop Science 4:530-537.

Shao, Y. H., W. X. Zhang, N. Eisenhauer, T. Liu, Y. M. Xiong, C. F. Liang, and S. L. Fu. 2017. Nitrogen deposition cancels out exotic earthworm effects on plant-feeding nematode communities. Journal of Animal Ecology 86:708-717.

Simon, S., B. Sauphanor, and P. E. Lauri. 2007. Control of fruit tree pests through manipulation of tree architecture. Pest Technology 1:33-37.

Sivapalan P. 1999. Pest management in tea. in Jain NK, ed. Global Advances in Tea Science. New Delhi: Aravali Books. 882 pp625-646.

Spiller, D. A., and A. A. Agrawal. 2003. Intense disturbance enhances plant susceptibility
to herbivory: natural and experimental evidence. Ecology 84:890-897.

Strauss, S. Y., and R. E. Irwin. 2004. Ecological and evolutionary consequences of multispecies animal-plant interactions. Annual Review of Ecology Evolution and Systematics 35:435-466.

Sun, S. G., W. S. Armbruster, and S. Q. Huang. 2016. Geographic consistency and variation in conflicting selection generated by pollinators and seed predators. Annals of botany 118:227-237.

Tworkoski, T., S. Miller, and R. Scorza. 2006. Relationship of pruning and growth morphology with hormone ratios in shoots of pillar and standard peach trees. Journal of Plant Growth Regulation 25:145-155.

Vile, D., B. Shipley, and E. Garnier. 2006. A structural equation model to integrate changes in functional strategies during old-field succession. Ecology 87:504-517.

Wen, X., M. Shi, L. Zhu, and D. Fu. 2006. Effect of pruning and ground treatments on the populations of Hylobitelus xiaoi Zhang, a new debarking weevil in slash pine plantations. Agricultural and forest entomology 8:263-265.

White, T. C. R. 1969. An Index to Measure Weather-Induced Stress of Trees Associated With Outbreaks of Psyllids in Australia. Ecology 50(5):905-909.

Zhang, D., M. Zhang, K. Li, J. Huo, Q. Yu, J. Lu, Q. Wang, Q. Li, and G. Chen. 2009. Analysis of climate change characteristics in Kaifeng in the last 50 years. Meteorological and Environmental Sciences 32:82-86. (In Chinese with English Abstract).

Appendix

Figure S1. Aphid number on the plants with a cubic canopy (hollow circles) and with a spherical canopy (solid circles) in the road green belt before the first pruning, which was conducted on May 5, 2015.

Figure S2. Flavonoid content of the leaves on the plants with a cubic (CU) and a spheroidal
canopy (SP) on the campus (CA), at the road green belt (RGB) and in the residential area (RA).

Figures

Figure 1

Photo of B. megistophylla Levl. with a cubic canopy (CU) and with a spheroidal canopy (SP).
Figure 2

Aphid abundance (a) and herbivore load (b) of plants with cubic (CU) and spheroidal canopies (SP) on the campus (CA), at the road green belt (RGB) and in the residential area (RA).
Figure 3

Fresh twig length (a), leaf dry mass content (LDMC, b) and inflorescence number (c) of plants with cubic (CU) and spheroidal canopies (SP) on the campus (CA), at the road green belt (RGB), and in the residential area (RA).
Linear regressions show the relationships between aphid abundance and leaf dry mass content (LDMC) across all samples (ALL, a, b, c) and samples of plants with cubic (CU, d, e, f) and spheroidal canopies (SP, g, h, i). The words above each column represent the different study sites; campus (CA), road green belt (RGB) and residential area (RA). All the data in this figure has been log-transformed.
Linear regressions show the relationships between aphid abundance and fresh twig length across all samples (ALL, a, b, c) and samples of cubic (CU, d, e, f) and spheroidal canopies (SP, g, h, i). The words above each column represent the different study sites; campus (CA), road green belt (RGB) and residential area (RA). All the data in this figure has been log-transformed.
Figure 6

Linear regressions show the relationships between aphid abundance and inflorescence number across all samples (ALL, a, b, c) and samples of cubic (CU, d, e, f) and spheroidal canopies (SP, g, h, i). The words above each column represent the different study sites; campus (CA), road green belt (RGB) and residential area (RA). All the data in this figure has been log-transformed.
The results of the final structural equation model (SEM) showing the causal relationships from pruning, leaf dry mass content, inflorescence number and twig length to aphid abundance. The blue line represents a positive effect and the red line represents a negative effect. The numbers above the arrows indicate path coefficients (*P < 0.05, **P < 0.01, ***P < 0.001). The values above the variables represent the proportion of variance explained for each variable. Model fit summary: $\chi^2 = 2.797$, $P = 0.094 > 0.05$, df = 1, TLI = 0.959, CFI = 0.996, RMSEAR = 0.175.

Supplementary Files

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