Resistance of aerobic microorganisms and soil enzyme response to soil contamination with Ekodiesel Ultra fuel

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Abstract This study determined the susceptibility of cultured soil microorganisms to the effects of Ekodiesel Ultra fuel (DO), to the enzymatic activity of soil and to soil contamination with PAHs. Studies into the effects of any type of oil products on reactions taking place in soil are necessary as particular fuels not only differ in the chemical composition of oil products but also in the composition of various fuel improvers and antimicrobial fuel additives. The subjects of the study included loamy sand and sandy loam which, in their natural state, have been classified into the soil subtype 3.1.1 Endocalcaric Cambisols. The soil was contaminated with the DO in amounts of 0, 5 and 10 cm3 kg−1. Differences were noted in the resistance of particular groups or genera of microorganisms to DO contamination in loamy sand (LS) and sandy loam (SL). In loamy sand and sandy loam, the most resistant microorganisms were oligotrophic spore-forming bacteria. The resistance of microorganisms to DO contamination was greater in LS than in SL. It decreased with the duration of exposure of microorganisms to the effects of DO. The factor of impact (IFDO) on the activity of particular enzymes varied. For dehydrogenases, urease, arylsulphatase and β-glucosidase, it had negative values, while for catalase, it had positive values and was close to 0 for acid phosphatase and alkaline phosphatase. However, in both soils, the noted index of biochemical activity of soil (BA) decreased with the increase in DO contamination. In addition, a positive correlation occurred between the degree of soil contamination and its PAH content.

Keywords Microbiota · Enzymatic activity · PAHs · Oil products · Degradation · Soil stability

Introduction

Progressive industrialisation and urbanisation contribute to the degradation of ever increasing areas, which leads to a reduction in biological diversity (Global Environment Outlook 2012). A significant role in the degradation of natural environments is performed by persistent organic pollutants, including polycyclic aromatic hydrocarbons (Angello et al. 2016). The European Commission, the European Economic and Social Committee and the Committee of the Regions (COM 2006) pay close attention to polycyclic aromatic hydrocarbons (PAHs) as compounds resulting in soil degradation.

The hazards to the soil environment and to human health are not only posed by exhaust emissions from diesel engines, including inter alia PAHs, but also by point contamination of soils with oil products due to failures, accidents, spills during reloading, etc. (Park and Park 2011; Ziolkowska and Wyszkowski 2010). The response of microorganisms and soil enzymes to the effects of particular oil products should be determined individually for particular types and brands of these products (Adam et al. 2017; Niepceron et al. 2013).
Not only do they differ in the natural chemical composition but also in the content of improvers, which offer anti-corrosive, cidal, demulsifying, and anti-foam properties and may modify the effects of particular oil products on the quality of soil (Comber et al. 2016; Ramkumar and Kirubakaran 2016). Therefore, a proper assessment of risks to soils contaminated with various hydrocarbons is necessary in order to understand and manage such ecosystems (Pinedo et al. 2012). Furthermore, identifying the metabolic response of soils to particular types of contaminants may facilitate their biotechnological remediation.

Many soil microorganisms are actively involved in restoring the biological balance of soils contaminated with oil products (Table 1). As for bacteria, the following genera are dominant: Bacillus (Bento et al. 2005; Fatima et al. 2015), Pseudomonas (Fatima et al. 2015), Staphylococcus (Moscoso et al. 2012; Silva et al. 2015), Acinetobacter (Chang et al. 2011; Mnif et al. 2015) and Paenibacillus (Lipińska et al. 2015), while as regards fungi, the following genera predominate: Aspergillus (Díaz-Ramírez et al. 2013; El-Hanafy et al. 2017) and Candida (Fan et al. 2014; Silva et al. 2015). These products can also be removed from soil by, inter alia, biostimulation of autochthonous microorganisms (Adam et al. 2017; Fan et al. 2014), bioaugmentation (Chang et al. 2011; Fan et al. 2014; Galiulina and Galiulina 2015; Semrany et al. 2012), phyto remediation (Agnelo et al. 2016; Sivitskaya and Wyszkowski 2013; Soleimani et al. 2010) and electro-bioremediation (Shrestha et al. 2010; Yuan et al. 2016; Wang et al. 2012). It is important, however, to observe the response of both cultured microorganisms (Kucharski and Jastrzębska 2005; Soleimani et al. 2010; Wyszkowska et al. 2015) to soil contamination and the enzymes of importance in terms of soil quality, i.e. those involved in respiration processes and transformations of carbon, nitrogen, phosphorus and sulphur (Baran et al. 2004; Lipińska et al. 2015; Wu et al. 2014; Wyszkowska et al. 2006). These are the agents which, in the early period following the contamination, indicate the degree of soil degradation (Perez-Leblic et al. 2010; Wu et al. 2014) and facilitate selecting the method of remediation. A good indicator of the biological condition is the measurement of the activity of dehydrogenases (Kaczynska et al. 2015; Lipińska et al. 2014a). The activity of these enzymes most accurately reflects the reactions taking place in soil as these are intracellular enzymes which means that they are closely related to the microbial activity of soil (Moeskops et al. 2010; Subhani et al. 2001). Urease activity is a rather good indicator as well. Admittedly, it is an extracellular enzyme less related to the condition of microorganisms, yet it is very sensitive to the effects of various xenobiotics (Lipińska et al. 2013; Zhan et al. 2010). An enzyme that is important in the assessment of the reactions taking place in soil is β-glucosidase, which is responsible for the final transformation of cellulose to glucose (Knight and Dick 2004), all the more so as transformations of organic carbon and nitrogen are relatively consistent with each other. The information on disturbances to transformations of organic phosphorus compounds is relatively accurately disclosed by the activity of acid phosphatase and alkaline phosphatase (Kucharski and Jastrzębska 2006; Wyszkowska et al. 2015), and the transformations of organic sulphur are reflected by the activity of arylsulphatase (Lipińska et al. 2014b; Vong et al. 2008). Particular attention should be paid to the relationship between the activity of alkaline phosphatase and the activity of acid phosphatase (Wyszowska and Wyszkowski 2010). However, in order to avoid randomness in estimating the enzymatic activity of soils, not only does the study assess the activity of particular enzymes but also use the index of activity of soil (BA), which reflects the total activity of all enzymes (Wyszkowska et al. 2013).

For the above reasons, it was decided to carry out a study aimed at the assessment of the response of soil microorganisms and enzymes to the contamination of soil with Ekodiesel Ultra fuel. A study of this type, conducted on this particular oil product, is innovative.

Material and methods

Characteristics of Ekodiesel Ultra fuel

The study tested Ekodiesel Ultra fuel of grade B, which in temperate climates is used from 16 April to 30 September. It is characterised by the following properties: density, 820–845 g dm⁻³; cetane number, min 51; cetane index, min 46; and content: PAHs—max 7% (m/m), solid impurities—max. 24 mg kg⁻¹, fatty acid methyl esters—max 7% (v/v), sulphur—max 10 mg kg⁻¹, manganese—max 2 mg dm⁻³ and water—max 200 mg kg⁻¹ (www.orlen.pl).

Soil

The contamination of soils with oil products reduces water capacity and hampers air exchange through filling soil pores and disturbing the C:N ratio. As the destabilisation of physical, chemical and biological properties of soil by diesel fuel is determined by the granulometric composition, the study used two soils significantly differing in the contents of clay, silt and sand (Table 2). For the study, soils typical of the early post-glacial landscape of north-eastern Poland were selected (Fig. 1). To this end, soils used for agricultural purposes, present at the Educational and Experimental Station in Tomaszkowo of the University of Warmia and Mazury in Olsztyn (NE Poland, 53.7167° N, 20.4167° E), were identified. The Educational and Experimental Station in Tomaszkowo is located in the Olsztyn Lakeland (Pojezierze Olsztyńskie), a physical and geographical meso-region which
is part of the Mazurian Lake District (*Pojezierze Mazurskie*). The soils that predominate there belong to Order 3: Brown earths of the soil type 3.1: Eutric Cambisols. Taking into account the classification in terms of grain size according to the World Reference Base of Soil Resources (2014), for further model study, a soil was selected of the subtype 3.1.1 Endocalcaric Cambisols. The soils were sampled from the tilled Ap horizon (a depth of 0–20 cm). In terms of grain size, they were loamy sand and sandy loam (Table 2).

**Experimental design**

Since oil products have great potential for accumulation in the soil environment and may lead to gradual degradation of soils,
it is particularly important that the changes to be expected under field conditions should be observed, in the first place, in specially designed model experiments conducted under controlled conditions. With this principle in mind, the authors of the study using Ekodiesel Ultra fuel carried out an experiment under monitored conditions, in a greenhouse, in polyethylene pots with a capacity of 3.5 dm$^3$, in 4 repetitions. In the experiment, the variable factors included the following:

- Type of soil formation: loamy sand and sandy loam
- The degree of the contamination of soil with Ekodiesel Ultra fuel in cubic centimeters per kilogram of soil DM: 0, 5 and 10
- The duration of incubation of soil samples, in days: 30 and 60

Prior to the establishment of the experiment, the selected and precisely pre-characterised soils were passed through a 2-mm sieve. The soil samples (each weighing 3 kg) were then mixed with Ekodiesel Ultra fuel and brought to a moisture content corresponding to 60% of the capillary water capacity. The moisture content of soil was monitored and maintained at a constant level throughout the duration of the experiment, i.e. for 60 days. In order to maintain the soil moisture content at a level of 60% of the capillary water capacity, soil in the pots was covered with perforated foil and weighed once a week.

### Table 2  Soil characteristics (granulometric composition, physiochemical and chemical properties, microorganisms CFU)

| Properties                                | Kind of soil   | Methods of determination |
|-------------------------------------------|----------------|--------------------------|
| Granulometric composition—% of fractions (d, mm) |                |                          |
| Sand 2.00 ≥ d > 0.05                      | Loamy sand (LS)| 75.56 ± 4.54             |
|                                           | Sandy loam (SL)| 47.92 ± 2.29             |
| Silt 0.05 ≥ d > 0.002                     |                | 22.92 ± 1.15             |
|                                           |                | 48.71 ± 1.70             |
| Clay d ≤ 0.002                           |                | 1.52 ± 0.08              |
|                                           |                | 3.37 ± 0.21              |
| pH$_{KCl}$                               |                | 6.7 ± 0.2                |
|                                           |                | 6.8 ± 0.2                |
| mmol(+) kg$^{-1}$                         |                |                          |
| Hydrolytic acidity (HAC)                 |                | 7.81 ± 0.41              |
|                                           |                | 5.22 ± 0.2               |
| Exchangeable base cations (EBC)          |                | 98.72 ± 4.89             |
|                                           |                | 131.41 ± 6.01            |
| Cation exchange capacity (CEC)           |                | 106.53 ± 5.18            |
|                                           |                | 136.63 ± 6.08            |
| % BS—base saturation                     |                | 92.67 ± 3.12             |
|                                           |                | 96.18 ± 3.38             |
| Content—g kg$^{-1}$                      |                |                          |
| Organic carbon (C$_{organic}$)           |                | 11.0 ± 0.52              |
|                                           |                | 9.92 ± 0.41              |
| Total nitrogen (N$_{total}$)             |                | 0.97 ± 0.05              |
|                                           |                | 1.14 ± 0.06              |
| Content mg kg$^{-1}$                     |                |                          |
| Available phosphorus (P$_{available}$)   |                | 44.92 ± 1.41             |
|                                           |                | 47.94 ± 1.24             |
| Exchangeable potassium (K$_{available}$) |                | 75.21 ± 3.87             |
|                                           |                | 69.13 ± 3.38             |
| Exchangeable magnesium (Mg$_{available}$) |                | 21.93 ± 1.13             |
|                                           |                | 12.74 ± 0.51             |
| Number of microorganisms—CFU 10$^n$ kg$^{-1}$ |                |                          |
| Oligotrophic bacteria                     |                | 14.53 ± 2.18             |
| Oligotrophic sporulating bacteria         |                | 5.72 ± 0.80              |
| Copiotrophic bacteria                     |                | 10.11 ± 1.36             |
| Copiotrophic sporulating bacteria         |                | 4.88 ± 0.49              |
| Azotobacter                               |                | 1.04 ± 0.12              |
| Arthrobacter                              |                | 10.24 ± 1.52             |
| Pseudomonas                               |                | 11.70 ± 1.76             |
| Ammonifying bacteria                      |                | 9.33 ± 1.52              |
| Nitrogen immobilisation bacteria          |                | 12.55 ± 1.07             |
| Cellulolytic bacteria                     |                | 17.88 ± 3.06             |
| Actinomycetes                             |                | 9.65 ± 1.02              |
| Fungi                                     |                | 4.29 ± 0.64              |

n—exponent: 8 for oligotrophic sporulating and copiotrophic sporulating bacteria; 7 for cellulolytic bacteria and Fungi; 4 for Azotobacter; 8 for Arthrobacter and Pseudomonas; 9 for other microorganisms. The same lowercase letters in the rows indicate homogeneous groups.
and any water losses were replenished. A separate series of soils was prepared for each date of testing.

The experiment was conducted in 2015, in the months June–July. The monthly average temperature was 15.5 °C in June and 17.9 °C in July. After 30 and 60 days of the experiment, soil samples were transported to a laboratory and kept at a temperature of 4 °C. Microbiological analyses were performed directly after taking the soil samples, and biochemical determinations were made on the following day. The physico-chemical properties of soil were determined in air-dry samples stored in the dark for 10 days.

**Determination of the soil microorganism number**

On days 30 and 60 of the experiment, soil samples were collected for analyses, and the next stage of the study was commenced in a laboratory. Microbiological analyses were first conducted using the soil culture dilution method. Soil samples (10 g) were weighed to a sterile physiological saline solution (90 cm³ of 0.85% NaCl) and shaken for 30 min at 120 rpm. Two appropriate dilutions of soil culture were inoculated into Petri dishes in parallel in four repetitions: oligotrophic bacteria (Olig), copiotrophic bacteria (Cop), nitrogen-immobilising bacteria (Im), ammonifying bacteria (Am) and actinobacteria (Act)—dil. of $10^{-5}$ and dil. of $10^{-6}$, Arthrobacter (Art), Pseudomonas (Ps)—dil. of $10^{-4}$ and dil. of $10^{-5}$, oligotrophic spore-forming bacteria (Oligp) and copiotrophic spore-forming bacteria (Copp)—dil. of $10^{-3}$ and dil. of $10^{-4}$, cellulolytic bacteria (Cel), and fungi (Fun)—dil. of $10^{-2}$ and dil. of $10^{-4}$ and Azotobacter spp. (Az)—dil. of $10^{0}$. Appropriate selective mediums were then introduced. The composition of microbiological mediums on which microorganisms were cultured is presented in Table 3. Microorganisms were cultured on Petri dishes at a temperature of 28 °C, within a period ranging from 2 (Azotobacter) to 21 days (oligotrophic bacteria). The spore forms of oligotrophic and copiotrophic bacteria were determined in the material which was pasteurised for 15 min at a temperature of 85 °C. The number of colony forming units (CFU) was determined using a colony counter.

**Determination of the activity of soil enzymes**

At the same time that the number of microorganisms was determined, i.e. on days 30 and 60 of the experiment in soil samples, from each repetition in three subsequent replications, the activity of dehydrogenases, catalase, urease, acid phosphatase, alkaline phosphatase, β-glucosidase and arylsulphatase...
was determined. The substrates used for the determination of the enzyme activity, as well as the units in which the activity of particular enzymes was expressed, are presented in Table 4. The activity of all enzymes, with the exception of catalase, was determined using a Perkin-Elmer Lambda 25 spectrophotometer (MA, USA). The activity of dehydrogenases was determined at a wavelength (λ) of 485 nm; the activity of urease, acid phosphatase and alkaline phosphatase at 410 nm; the activity of β-glucosidase at 400 nm; and the activity of arylsulphatase at 420 nm. The activity of catalase was determined based on the reaction of hydrogen peroxide decomposition using potassium permanganate.

### Table 3

| Medium for determination of number of soil microorganisms | Microorganisms | Medium | References |
|---------------------------------------------------------|----------------|--------|------------|
| Peptone 0.10 g; meat extract 0.10 g; NaCl 0.05 g; agar 10.0 g; H₂O 1.0 dm³; pH 7.0–7.2 | Olig—oligotrophic bacteria | [Ohta and Hattori (1983)] |
| Peptone 0.10 g; NaCl 0.3 g; FeSO₄·7H₂O 0.005 g; MnSO₄·7H₂O 0.005 g; CaCO₃ 15.0 g; agar 7.0 g; H₂O 1.0 dm³; pH 7.0–7.2 | Olig—oligotrophic spore-forming bacteria | [Fenglerowa (1965)] |
| Peptone 0.25 g; K₂HPO₄ 1.0 g; MgSO₄·7H₂O 0.25 g; glycerol 10 cm³; agar 14.0 g; H₂O 1.0 dm³; pH 7.0 | Cop—copiotrophic bacteria | [Mulder and Antheumisse (1963)] |
| Peptone 10 g; K₂HPO₄ 1.5 g; MgSO₄·7H₂O 1.5 g; glycerol 10 cm³; agar 14.0 g; H₂O 1.0 dm³; pH 7.0–7.2 | Cop—copiotrophic spore-forming bacteria | [Mulder and Antheumisse (1963)] |
| K₂HPO₄ 1.5 g; MgSO₄·7H₂O 0.3 g; NaCl 0.3 g; FeSO₄·7H₂O 0.005 g; MnSO₄·7H₂O 0.005 g; CaCO₃ 15.0 g; agar 7.0 g; H₂O 1.0 dm³; pH 7.0–7.2 | Az—Azotobacter | [Ohta and Hattori (1983)] |
| CaH₂PO₄ 0.25 g; K₂HPO₄ 1.0 g; MgSO₄·7H₂O 0.25 g; glycerol 10 cm³; agar 14.0 g; H₂O 1.0 dm³; pH 7.0–7.2 | Art—Arthrobacter | [Mulder and Antheumisse (1963)] |
| Peptone 0.2 g; NaCl 0.2 g; MnSO₄ 0.005 g; FeSO₄ 0.005 g; NH₄(NO₃) 0.025 g; H₂O 1.0 dm³; pH 7.0–7.2 | Ps—Pseudomonas | [Mulder and Antheumisse (1963)] |
| Peptone 20 g; K₂HPO₄ 1.5 g; MgSO₄·7H₂O 1.5 g; agar 14.0 g; H₂O 1.0 dm³; pH 7.0–7.2 | Am—ammonifying bacteria | [Winogradski (1953)] |
| Peptone 0.1 g; NaCl 0.2 g; MnSO₄ 0.005 g; FeSO₄ 0.005 g; (NH₄)₂Fe(SO₄)₂·6H₂O 0.15 g; agar 14 g; H₂O 1.0 dm³; pH 7.0–7.2 | Im—nitrogen immobilisation bacteria | [Winogradski (1953), our modification—instead of (NH₄)₂SO₄ used NH₄NO₃] |
| Peptone 0.1 g; NaCl 0.2 g; MnSO₄ 0.005 g; FeSO₄ 0.005 g; (NH₄)₂Fe(SO₄)₂·6H₂O 0.05 g; agar 14 g; H₂O 1.0 dm³; pH 7.0–7.2 | Cel—cellulolytic bacteria | [Winogradski (1953)] |
| Soluble starch 25 g; casein 0.3 g; KNO₃ 2.0 g; NaCl 2.0 g; K₂HPO₄ 2.0 g; MgSO₄·7H₂O 0.05 g; CaCO₃ 0.02 g; FeSO₄ 0.01 g; agar 20.0 g; H₂O 1.0 dm³; 5.0 cm³ aqueous solution of nystatin 0.05%; 5.0 cm³ aqueous solution of actidione 0.05%; pH 7.0 | Act—actinomycetes | [Parkinson et al. (1971)] |
| Peptone 0.1 g; glucose 10 g; MgSO₄·7H₂O 0.5 g; agar 20.0 g; H₂O 1.0 dm³; 3.3 cm³ aqueous solution of bengal rose 1%; 25 cm³ aqueous solution of aureomycin 0.01%; pH 5.9 | Fun—fungi | [Martin (1950)] |

### Table 4

| Enzyme | Substrate | Product/unit | References |
|--------|-----------|-------------|------------|
| Deh—dehydrogenases (EC 1.1) | 2,3,5-Triphenyl tetrazolium chloride (TTC) | Triphenyl formazan (TFF), μmol kg⁻¹ DM of soil h⁻¹ | Øhlinger (1996) |
| Cat—catalase (EC 1.11.1.6) | H₂O₂—aqueous solution | O₂, mol kg⁻¹ DM of soil h⁻¹ | Alef and Nannipieri (1998) |
| Ure—urease (EC 3.5.1.5) | Urea—aqueous solution | N-NH₄, mmol kg⁻¹ DM of soil h⁻¹ | Alef and Nannipieri (1998) |
| Glu—β-glucosidase (EC 3.2.1.21) | 4-Nitrophenyl-β-D-glucopyranoside (PNG) | 4-Nitrophenol (PN), mmol kg⁻¹ DM of soil h⁻¹ | Alef and Nannipieri (1998) |
| Pac—acid phosphatase (EC 3.1.3.2) | Disodium 4-nitrophenyl phosphate hexahydrate (PNP) | 4-Nitrophenol (PN), mmol kg⁻¹ DM of soil h⁻¹ | Alef and Nannipieri (1998) |
| Pal—alkaline phosphatase (EC 3.1.3.1) | Disodium 4-nitrophenyl phosphate hexahydrate (PNP) | 4-Nitrophenol (PN), mmol kg⁻¹ DM of soil h⁻¹ | Alef and Nannipieri (1998) |
| Aryl—arylosulphatase (EC 3.1.6.1) | Potassium-4-nitrophenylsulfate (PNS) | 4-Nitrophenol (PN), mmol kg⁻¹ DM of soil h⁻¹ | Alef and Nannipieri (1998) |
**Determination of PAH content**

In soil samples (loamy sand and sandy loam) contaminated with Ekodiesel Ultra fuel, after 60 days, the contents of 9 PAHs were determined, i.e. naphthalene (NAP), phenanthrene (PHE), anthracene (ANT), fluoranthene (FTH), benzo(α)-antracene (BaA), chrysene (CHR), benzo(α)fluoranthene (BaF), benzo(α)pyrene (BaP) and benzo(ghi)perylene (BghiP). The reference was the determination of PAH contents in samples of non-contaminated soil. The content of polycyclic aromatic hydrocarbons was determined using an Agilent 7890A gas chromatograph coupled with a Agilent 5975C mass spectrometer equipped with an EI/CI ion source. The hydrocarbons were extracted in accordance with the standard ISO 18287 (2006).

**Determination of physicochemical and chemical properties of the soil**

Prior to the establishment of the experiment, and on day 60 of the experiment, soil samples were taken (from each pot) and then dried and passed through a 2-mm sieve. In the prepared soil, the following were then determined: granulometric composition of soil, pH value of soil, hydrolytic acidity (HAC) and exchangeable base cations (EBC). Based on the HAC and EBC values, the sorption capacity of soil (CEC) and the degree of soil saturation with basic elements (BS) were determined. The following formulas were applied: CEC = EBC + HAC; BS = (EBC/CEC)·100. The contents of total nitrogen, organic carbon (Corg), available phosphorus, potassium and magnesium were also determined. The methods applied to determine the physicochemical and chemical properties are presented in Table 2.

**Methodology of calculations**

The varied effects of Ekodiesel Ultra fuel on the soil microbiota are presented using the index of resistance of microorganisms (RS), the factor of impact of diesel fuel (IFDO) and the index of biochemical activity of soil (BA). Taking into account the number of 12 groups of microorganisms, their resistance to the effects of diesel fuel was calculated. To this end, the formula provided by Orwin and Wardle (2004) was applied:

\[
RS = 1 - \frac{2|D_0|}{C_0 + |D_0|}
\]  

The RS index takes values from −1 to +1. RS with a value of 1 means full resistance; 0—a 100% decrease, or a 100% increase of the tested characteristic; negative values—an increase by more than 100%, the more the RS is close to −1, the greater the increase—by more than 100%.

The factor of impact of diesel fuel (IFDO) on the activity of soil enzymes was calculated according to formula no. 2 and the index of biochemical activity of soil (BA) was calculated based on formula no. 3 (Wyszkowska et al. 2013).

\[
IFDO = \frac{P_D - C_0}{C_0}
\]  

**Statistical analysis**

The study results were analysed statistically. In the interpretation of the effects of Ekodiesel Ultra fuel on shaping the number of microorganisms and the enzymatic activity, it was helpful to determine the percentages of particular independent variables in shaping the dependent variables. To this end, an analysis of the measure of the effect \(\eta^2\) was applied and carried out using the variance analysis ANOVA method. The homogeneous groups were counted using Tukey’s test at \(P = 0.01\). In the assessment of the effects of Ekodiesel Ultra fuel on microbiota and soil enzymes, cluster analysis (CA) was applied. In order to estimate the distance between clusters, variance analysis was applied. The distance between clusters was measured using Euclidean distance by Ward’s method. The results were also analysed using a principal component analysis (PCA) test, and correlation coefficients were calculated between the PCA.
results. In the statistical processing of results, a STATISTICA 12.5 package (StatSoft, Inc. 2015) was used.

**Results**

**Soil microorganisms**

Initial CFU of microorganisms in soil were included in Table 2. Of the three tested factors, i.e. the type of soil, the degree of contamination with Ekodiesel Ultra fuel and the duration of incubation, the number of soil microorganisms was most significantly affected by the first two factors (Fig. 2). Their participation in shaping the number certainly varied for particular groups and genera of microorganisms. The type of soil determined, to the greatest extent, the number of the following bacteria: *Azotobacter* sp. (77%), *Arthrobacter* sp. (56%), copiotrophic spore-forming bacteria (55%), cellulolytic bacteria (49%), nitrogen-immobilising bacteria (31%), *Pseudomonas* sp. (29%), actinobacteria (25%) and total copiotrophic bacteria (24%). The type of soil affected other groups of microorganisms to a much lesser degree. The impact of this factor ranged from 2% (Olig sp) to 12% (Olig). In turn, the contamination of soil with Ekodiesel Ultra fuel had the greatest effect on the number of the following: ammonifying bacteria (81%), actinobacteria (53%), oligotrophic bacteria (41%), nitrogen-immobilising bacteria (36%), copiotrophic bacteria (28%) and fungi (48%). The number of other groups of microorganisms was to a much lesser degree determined by the contamination with Ekodiesel Ultra fuel, and the percentage of this contaminant in the structure of the number ranged from 2% (cellulolytic bacteria) to 11% (*Pseudomonas* sp.).

Therefore, the response of microorganisms to the contamination of soil with Ekodiesel Ultra fuel varied (Fig. 3). A significant correlation occurred between the number of *Arthrobacter* sp. and *Azotobacter* sp. ($r = 0.52$, $P = 0.001$), actinobacteria and oligotrophic bacteria ($r = 0.64$, $P < 0.001$), actinobacteria and nitrogen-immobilising bacteria ($r = 0.88$, $P < 0.001$), oligotrophic and nitrogen-immobilising bacteria ($r = 0.70$, $P < 0.001$), total copiotrophic bacteria and ammonifiers bacteria ($r = 0.37$, $P = 0.026$) and cellulolytic and copiotrophic spore-forming bacteria ($r = 0.62$, $P < 0.001$). Such a relationship is not only proven by the data presented using the PCA method but also by the similarity of responses of particular groups of microorganisms, documented by Ward’s cluster method (Fig. 4).

The greatest increase in the number (Fig. 3) of actinobacteria, as well as oligotrophic and nitrogen-immobilising bacteria, was observed in sandy loam contaminated with 10 cm$^3$ of DO, irrespective of the duration of soil incubation, while the greatest increase in the number of copiotrophic bacteria and their spore forms, ammonifying and cellulolytic bacteria, was observed in loamy sand contaminated with 10 cm$^3$ of DO on both dates of the study and in loamy sand contaminated with 5 cm$^3$ of DO on the second date of the study. In the mentioned objects, a negative response of bacteria of the *Arthrobacter* and *Azotobacter* genera to the contamination with DO occurred. The response of fungi to contamination with DO was unequivocally negative in both soils, irrespective of the date of testing.

In general, the lower the average resistance of microorganisms to the effects of DO was, the greater the contamination with this pollutant was, i.e. it was lower in sandy loam than in loamy sand; moreover, it was related to the smallest extent to the date of testing (Fig. 5). There is certainly a significant difference between the resistances of particular groups. In loamy sand, the most resistant bacteria (Fig. 6) were oligotrophic spore-forming bacteria (RS = 0.63–0.95), while the least resistant were bacteria of the *Azotobacter* genus (RS = 0.06–0.34). In sandy loam, the most resistant bacteria were oligotrophic spore-forming bacteria (RS = 0.71–0.95), while the least resistant were bacteria of the nitrogen-immobilising...
bacteria (RS = −0.42 to −0.78). With time, the resistance of both spore-forming groups decreased.

Physicochemical properties

The impact of fuel contamination on physicochemical properties was minimal (Table 5). In both loamy sand and sandy loam, the amount of organic carbon in contaminated objects increased. The pool of total nitrogen and available potassium did not change, while the contents of available phosphorus and magnesium only changed in loamy sand. The soil pH and hydrolytic acidity were stable. In loamy sand, under the effect of Ekodiesel Ultra fuel, the total content of exchangeable cations and the exchange capacity increased, while the degree of soil saturation with alkaline cations of both soils remained unchanged.

Soil enzymes

The impact degree of the tested independent variables of the activity of soil enzymes varied, similar to the impact degree on the activity of soil microorganisms (Fig. 2). The Ekodiesel Ultra fuel had the strongest effect on the activity of dehydrogenases, catalase, urease and arylsulphatase and the weakest effect on the activity of acid phosphatase, β-glucosidase and alkaline phosphatase. The activity of dehydrogenases was...
determined in 37% by the contamination of soil with Ekodiesel Ultra fuel (catalase in 62%; urease in 36%; arylsulphatase in 25%; acid phosphatase in 1.5%; β-glucosidase in 4%; and alkaline phosphatase in 10%). The percentage participation of the type of soil in shaping the activity of enzymes ranged from 14% (urease) to 57% (alkaline phosphatase), whereas the duration of incubation ranged from 3 to 7% (dehydrogenases, catalase, urease, arylsulphatase), 22–24% (alkaline phosphatase, β-glucosidase) and to 66% (acid phosphatase).

A PCA demonstrated that the response of dehydrogenases to Ekodiesel Ultra fuel was inverse to the response of catalase (Fig. 3). Dehydrogenases, acid phosphatase and β-glucosidase were strongly correlated with one another, while the other group comprised urease, arylsulphatase and alkaline phosphatase. This results from both the data shown in Fig. 3 and from the data shown in Fig. 4, which presents the relationship between the activities of enzymes using Ward’s cluster analysis method.

Catalase was the only enzyme that was stimulated by Ekodiesel Ultra fuel. Such an effect was noted in both soils, on both dates of testing. The impact factor (IF_{DO}) had positive values for catalase, both in loamy sand and in sandy loam and ranged from 0.07 to 0.93 in the first soil and from 0.16 to 0.59 in the second soil (Fig. 7). Negative values of IF_{DO} were noted for the remaining enzymes. The greatest inhibition of the...
activity occurred for urease \((\text{IF DO} = -0.14 \text{ to } -0.81)\), dehydrogenases \((\text{IF DO} = -0.01 \text{ to } -0.62)\) and arylsulphatase \((\text{IF DO} = -0.07 \text{ to } -0.37)\). The activity of other enzymes was inhibited to a much lesser degree. Such tendencies persisted during both parts of the study.

Irrespective of the stimulating effect of the Ekodiesel Ultra fuel on catalase, the quality of loamy sand and sandy loam, measured using the BA index, significantly decreased (Fig. 8). The impact factor \((\text{IF DO})\) on the BA value was negative on both dates of testing, and it became increasingly negative with greater soil contamination (Fig. 5).

**PAHs**

The contamination of soil with Ekodiesel Ultra fuel increased the PAH content of both tested soils (Fig. 9). In loamy sand, under the effect of Ekodiesel Ultra fuel at an amount of \(5 \text{ cm}^3 \text{ kg}^{-1}\), a 94% increase in the total PAH content occurred, while under the effect of an amount of \(10 \text{ cm}^3 \text{ kg}^{-1}\), an increase by up to 252% was observed; in sandy loam contaminated with the lower amount of Ekodiesel Ultra fuel, an increase in PAH content by 65% occurred, while for the higher amount, it was 214%. The concentration of naphthalene under the effect of \(10 \text{ cm}^3\) of Ekodiesel Ultra fuel increased in loamy sand by 33 times, and in sandy loam, by 28 times; for benz\((a)\)anthracene, by six and five times, respectively; for anthracene, by four and two times; for benzo\((a)\)pyrene, by 1.9 and 1.6; for fluoranthene, by four times in both soils; and for benzo\((a)\)fluoranthene, by three times, also in both soils. Despite the significant increase in aromatic hydrocarbon content of soils resulting from the contamination with Ekodiesel Ultra fuel, the concentration of each hydrocarbon did not exceed the acceptable standards. The soils contaminated with Ekodiesel Ultra fuel contained the most fluoranthene \((86 \mu \text{g kg}^{-1} \text{ DM})\) and the least benzo\((a)\)fluoranthene \((9 \mu \text{g kg}^{-1} \text{ DM})\).

**Discussion**

**Effect of diesel oil and soil physicochemical properties on soil microorganisms**

The response of microorganisms to soil contamination with Ekodiesel Ultra fuel is not clear. Not only was it determined by the physiological group of microorganisms and by their genus but also by the type of soil. The type of soil produces effects, not only as a specific living place, but also acts on microorganisms as a protective buffer (Griffiths and Philippot 2013).
Table 5. The effect of diesel oil on the physicochemical properties of the soil

| Soil Type       | Corg | Nog | PK | C:N | P:N | EBC | CEC | BS [%] | 
|-----------------|------|-----|----|-----|-----|-----|-----|--------|
|                 | Loamy sand (LS) | 0   | 5  | 10  |      |     |     |        |
|                 | Sandy loam (SL) | 0   | 5  | 10  |      |     |     |        |

In our study, Ekodiesel Ultra fuel changes values of EBC and CEC only in loamy sand. Sandy clay, being a more buffered soil, was more resistant to the impact of Ekodiesel Ultra fuel and the analysed soil properties maintained similar values, regardless of the contamination. Highly positive correlation with the soil sorption capacity was demonstrated for the bacteria Azotobacter \( r = 0.79, P = 0.002 \) and Arthrobacter \( r = 0.61, P = 0.035 \), whereas cellulolytic bacteria were highly negatively correlated with this characteristic \( r = -0.72, P = 0.009 \). Negative correlations were determined between the HAC value and the following bacteria: Azotobacter \( r = -0.86, P < 0.001 \), nitrogen-fixing \( r = -0.59, P = 0.045 \) and Arthrobacter \( r = -0.58, P = 0.045 \). In soil polluted with Ekodiesel Ultra fuel, significant positive correlations between the C:N ratio and groups of microorganisms were determined only for sporulating copiotrophic bacteria \( r = 0.61, P = 0.035 \) and ammonifying bacteria \( r = 0.62, P = 0.032 \), while significant negative correlations were detected between the C:N ratio and Azotobacter \( r = -0.62, P = 0.032 \). There were positive interactions between the C:P ratio and copiotrophic bacteria \( r = 0.58, P = 0.050 \), copiotrophic sporulating bacteria \( r = 0.60, P = 0.041 \) and ammonifying bacteria \( r = 0.69, P = 0.014 \). Also, a positive correlation was observed between the N:P ratio and oligotrophic bacteria \( r = 0.79, P = 0.002 \), nitrogen-fixing bacteria \( r = 0.80, P = 0.002 \), Arthrobacter \( r = 0.74, P = 0.006 \), Pseudomonas \( r = 0.60, P = 0.039 \) and Actinomyces \( r = 0.79, P = 0.002 \). Noteworthy is a very weak correlation between the analysed microorganisms and soil sorption capacity, which is a factor which contributes the most to the soil’s availability of nutrients to microorganisms (Canbolat et al. 2006). The most probable reason was the interference of the contamination with the physical properties of the soil (Caravaca and Rodán 2003; Semrany et al. 2012).

This accounts for the differences between the number of microorganisms and their diversity in loamy sand and in sandy loam. In addition, the varied response of autochthonous soil microorganisms to the tested diesel fuel might have resulted from the succession of microorganisms (Vazquez et al. 2013; Wu et al. 2014). New populations of bacteria and fungi may appear, which, due to intense metabolism, are capable of degrading PAHs and thus strongly affecting the soil microbiota (Semrany et al. 2012; Wyszkowska et al. 2015). According to Yemashova et al. (2007), the count of bacteria able to use petroleum carbohydrates as a source of energy in unpolluted soil is \( 10^{2}−10^{3} \) in 1 g of soil, while in polluted soil, it equals \( 10^{6}−5 \times 10^{7} \) CFU g\(^{-1}\). The dominant bacteria which effectively participate in decomposition of crude oil are as follows: Acinetobacter, Bacillus, Pseudomonas and Staphylococcus (Fatima et al. 2015; Moscoso et al. 2012; Mnif et al. 2015), and the fungi with such an ability include Aspergillus and Candida (El-Hanafy et al. 2017; Fan et al. 2014; Silva et al. 2015). In general, changes in the communities of
microorganisms are periodic, which is evidenced by fluctuations in the structure of r-strategy and K-strategy microorganisms (De Leij et al. 1993; Dorodnikov et al. 2009; Ernebjerga and Kishony 2012). In soils contaminated with petroleum products, fast-growing (r-strategy) microorganisms begin to dominate. After successful soil remediation, it can be expected that the state of equilibrium between these assemblages will be restored (De Leij et al. 1993; Dorodnikov et al. 2009; Ernebjerga and Kishony 2012). A stimulating effect of diesel oil on the number of copiotrophic, copiotrophic spore-forming and oligotrophic microorganisms, as well as actinobacteria, and a negative effect on cellulolytic bacteria and Azotobacter were observed by Wyszkowska and Kucharski (2005). Changes in the species composition of microorganism in the soil environment in terms of the fertility of soil are an adverse phenomenon. Changes to the number of microorganisms, which were observed in the authors’ own study in soil contaminated with Ekodiesel Ultra fuel, did not translate, however, into the resistance of microorganisms to the effects of DO, since both the stimulation and the inhibition of the proliferation of microorganisms by DO indicate the absence of their resistance. Even though the identification of the relationships between the indices describing the resistance (RS) of microorganisms is complicated and determined by numerous factors (Orwin and Wardle 2004), the determination in a study of the index of resistance enabled drawing objective conclusions on the stability of soil subjected to the effect of DO. The relatively low resistance of the tested microorganisms persisted for the entire duration of the study. It was lower with greater contamination with DO, yet definitely higher in loamy sand than in

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**Fig. 7** Impact factor (IFDO) of Ekodiesel Ultra fuel (DO) on activity of soil enzymes depending on the kind of soil. Explanations are provided in Tables 3 and 4. Identical lowercase letters for soil enzymes are assigned to homogeneous groups ($P < 0.01$)

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**Fig. 8** Effect of Ekodiesel Ultra fuel (DO) on biochemical activity of soil (BA)
sandy loam. This is logical since differences in the abundance and bioavailability of elements are determined by, inter alia, the contents of clay fraction and the silt fraction in soil (Wyszkowska and Kucharski 2005). In the authors’ own study, the contents of silt and clay in sandy loam were more than two times greater than in loamy sand (Table 2). The change in oxygenation of soils resulting from the effect of DO is also of significance. The latter factor probably determined the clearly negative effects of DO on bacteria of the Azotobacter genus and on fungi. For Azotobacter, many aromatic hydrocarbons are a source of carbon and energy, while for PAHs contained in DO, such an effect was not observed. According to Semrany et al. (2012), partial or total inhibition of aerobic microorganisms’ metabolism results from the poor solubility of these substances in water.

Oil products are contaminants hazardous to the soil environment, which cause long-term changes to the reactions taking place in soil. They also deteriorate their physical and chemical properties (Caravaca and Rodán 2003; Semrany et al. 2012). While penetrating into the soil profile, they cause changes to the oxide reduction potential of soils (Griffiths and Philippot 2013), as even the most stable substances in the soil environment can be metabolised by microorganisms (Adam et al. 2017; Fan et al. 2014; Fatima et al. 2015; El-Hanafy et al. 2017). Even though oil-derived hydrocarbons may be degraded by bacteria, fungi, yeasts and algae, it is bacteria that perform an essential role in the transformation of PAHs (Table 1). Biodegradation of PAHs in the soil environment is more often carried out by a consortium of microorganisms than by a particular species (Jung et al. 2016; Yergeau et al. 2012). Therefore, it is very important to identify the response of autochthonous microorganisms occurring in natural soil ecosystems on particular contaminants.

**Effect of diesel oil and soil physicochemical properties on soil enzymes**

The stability of soil ecosystems is very well described by enzymatic activity, which is a sensitive indicator of changes in soil quality taking place in real time (Burns et al. 2013). The most important enzymes are those involved in the element cycle and in the degradation of chemical substances. The most reliable indicators used for estimating the quality of soils are considered to be dehydrogenases (the opinion of 28% of researchers), phosphatase (28%), β-glucosidase (16%) and urease (11%) (Gil-Sotres et al. 2005). Also in our own study, Ekodiesel Ultra fuel, while disturbing the metabolic profile of soil, changed the activity of enzymes involved in the processes of transformation of carbon, nitrogen, phosphorus and sulphur. The change in biochemical properties caused by the tested contaminant was reflected in the impact factors of diesel fuel on particular enzymes. The identification of these indicators enabled an objective conclusion as to whether the tested ecosystem is stable and capable of maintaining proper
homeostasis. The factors of DO impact on the activity of all soil enzymes with the exception of catalase had negative values. These indices highlighted the disturbances to the metabolic profile of soil. On average, irrespective of the granulometric composition of soil or the duration of the effect of Ekodiesel Ultra oil on the reactions taking place in the soil, the enzymes may be ranked, in terms of their sensitivity to Ekodiesel Ultra, as follows (from the most to the least sensitive): urease > dehydrogenases > arylsulphatase > alkaline sulphatase > β-glucosidase > acid phosphatase. A relatively high resistance of enzymes to diesel fuel was also obtained by Kucharski and Jastrzębska (2005) and by Wyszkowska and Kucharski (2005). The changes to the biochemical activity of soils observed in this study were a response to biotic stress caused by contamination of soil with Ekodiesel ULTRA fuel. Certainly, the scale of changes to the biochemical activity of soils is determined by the type of oil product (Kucharski and Jastrzębska 2006; Wu et al. 2014; Wyszkowska et al. 2006). This is logical since these substances are a mixture of organic compounds with low bioavailability. The adverse effect of diesel fuel on the biochemical activity of soil is not a rare phenomenon. Serrano et al. (2009) also demonstrated the negative correlation between the contamination of soil with diesel fuel and the activity of dehydrogenase, arylsulphatase, protease, phosphatases and urease. On the other hand, β-glucosidase exhibited no significant response to the tested product. Wyszkowska et al. (2015) also found that β-glucosidase was the least resistant enzyme of the seven tested soil enzymes. In turn, according to Galilin et al. (2012), the decomposition of oil-derived hydrocarbons increases in parallel with the increase in the activity of dehydrogenases which are directly involved in this process. The stimulating effect of diesel fuel and fuel oil on the activity of soil enzymes is indicated by the results of studies by Kucharski and Jastrzębska (2006) and by Wyszkowska et al. (2002a). Studies conducted by Wyszkowska et al. (2006) in soil contaminated with diesel fuel in doses ranging from 3 to 24 cm³ kg⁻¹ of soil DM indicated an increase in the activity of dehydrogenases, urease and alkaline phosphatase. Wang et al. (2010) found that a mixture of hydrocarbons is particularly hazardous to the health of soil and such a mixture is present in oil products. According to the above-mentioned authors, hydrocarbons may both stimulate and inhibit the activity of phosphatases and catalase, although they are completely destructive to dehydrogenases, invertase and urease. In this study, the authors, in order to assess the quality of soil contaminated with Ekodiesel Ultra fuel as well as the activity of enzymes of the oxidoreductase and hydrolase classes, applied the index of biochemical activity (BA), which combines the activities of particular enzymes and more accurately reflects their responses to the tested contaminant. This index clearly demonstrated that disturbances to the soil metabolic profile were significant. Changes to the biochemical activity of soil result from the degree of degradation of Ekodiesel Ultra fuel, which was not only determined by the duration of DO retention in soil but also by the physicochemical properties of soil. In our study, it was loamy sand with pHKCl = 6.70 and carbon content of 11.05 g C kg⁻¹ of soil DM and sandy loam with pHKCl = 6.80 and carbon content of 10.25 g C kg⁻¹ of soil DM. Therefore, these were soils providing good conditions for the development of soil microbiota. This is important because certain microorganisms could use Ekodiesel Ultra fuel as a source of nutrients which, as a consequence, could affect the biosynthesis of enzymes by inducing or repressing phenomena. In this study, the activity of two enzymes, namely alkaline phosphatase (r = −0.67, P = 0.017) and arylsulphatase (r = −0.75, P = 0.005), was significantly negatively correlated with the N:P ratio. The same two enzymes were significantly negatively correlated with the C:P value (the activity of alkaline phosphatase at r = −0.64, P = 0.025, and the activity of arylsulphatase at r = −0.74, P = 0.006). The N:P ratio was positively correlated only with the activity of dehydrogenases (r = −0.66, P = 0.020) and catalase (r = 0.81, P = 0.001). The activity of catalase was positively correlated with the CEC value (r = 0.66, P = 0.019), the same as the activity of alkaline phosphatase (r = 0.71, P = 0.009) and arylsulphatase (r = 0.63; P = 0.029). A significant positive correlation occurred between pH and the activity of alkaline phosphatase (r = −0.71, P = 0.010) and arylsulphatase (r = −0.74, P = 0.006), which resulted in negative coefficients of correlation between HAC and the activity of alkaline phosphatase (r = −0.78, P = 0.003) and arylsulphatase (r = −0.73, P = 0.007). To recapitulate, there was an unequivocal interaction between the activity of enzymes and the physicochemical properties of soil, which arose from the unstable conditions created in soil polluted with Ekodiesel Ultra fuel (Griffiths and Philippot 2013; Semrany et al. 2012; Wyszkowska and Kucharski 2005).

An increase in the soil content of PAHs is one of the factors contributing to the impact of diesel oil on soil enzymes. This study had demonstrated only a significant positive correlation between the activity of catalase and the PAH content, as well as a negative correlation between these hydrocarbons and the activity of dehydrogenases, urease and arylsulphatase. The contents of all tested PAHs were significantly higher in both tested soils contaminated with Ekodiesel Ultra fuel than in the non-contaminated soil. A high percentage of 2- and 3-ring PAHs in the total PAH content was noted. Similar relationships were also demonstrated in studies by Nganje et al. (2007) and by Wyszkowska et al. (2015). According to Baran and Oleśczuk (2001), phenanthrene and fluoranthene, i.e. 2–4-ring PAHs, account for 70% of the total contamination with PAHs. According to Wyszkowski and Ziółkowska (2013), the different PAH contents of soil contaminated with oil products depend on the type of contaminant, physicochemical properties of soil (principally organic substance content) and crop species. The effects of Ekodiesel Ultra fuel on PAH content were stronger in soil on which spring barley, yellow lupine and maize were cultivated, while the effects of petrol on
PAH content were stronger in soil on which spring rape and oats were cultivated. Despite the significant increase in the aromatic hydrocarbon content of soils contaminated with Ekodiesel Ultra fuel and petrol, the contents of particular PAHs usually did not exceed 100 μg kg⁻¹ of soil DM. Similar results were obtained in the authors’ own study.

Conclusions

Disturbances in the reactions taking place in soil should be considered in terms of the possibility for restoring balance within the ecosystem. Therefore, the conducted experiment is an important link in the cycle of studies into the quality of the environment in which we live. The results presented in this study clearly indicate that the determination of the effects of oil products on the biochemical activity of soil is not a simple correlation. Significant differences were noted in the resistance of particular groups of microorganisms to the contamination of Endocalcaric Cambisol with Ekodiesel Ultra fuel. In loamy sand and sandy loam, the most resistant microorganisms were oligotrophic spore-forming bacteria. The resistance of microorganisms to contamination with Ekodiesel Ultra fuel was higher in loamy sand than in sandy loam. It decreased with the duration of exposure of microorganisms to the effects of Ekodiesel Ultra fuel. The effects of Ekodiesel Ultra fuel on the activity of particular enzymes varied. For dehydrogenases, urease, arylsulphatase and β-glucosidase, the impact factor IF_{DO} had negative values; for catalase, it had positive values, while for acid phosphatase and alkaline phosphatase, it was close to 0. A good measure of the effects of Ekodiesel Ultra fuel on the biochemical activity of soil was the BA index. In addition, there was a correlation between the degree of soil contamination and its PAH content. Additional data for the assessment of the condition of soils polluted with Ekodiesel Ultra fuel originated from the determination of interactions between the physicochemical properties of these soils and the soil microbiological activity. Therefore, the combination of microbiological and biochemical indicators with the parallel determination of physicochemical properties provides an opportunity to carry out a comprehensive analysis of soil quality under the pressure of oil-derived hydrocarbons.

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