Chapter 6
Delayed Recovery from Declines in the Population Densities and Species Richness of Intertidal Invertebrates Near Fukushima Daiichi Nuclear Power Plant

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Abstract  In June 2014, May and June 2015, and June 2016, we conducted quantitative quadrat surveys of sessile invertebrates at seven intertidal sites in Ibaraki, Fukushima, and Miyagi Prefectures, including the sites near Fukushima Daiichi Nuclear Power Plant (FNPP), to check whether species richness, population densities, and biomass had recovered from declines after the 2011 Tohoku earthquake, tsunami and nuclear disaster. Additionally, in April, July, August and September from 2012 to 2017, we monitored the population density and spawning behavior of rock shells (Thais clavigera) in the field near FNPP. Increases in species richness and population densities in the intertidal zone near FNPP were not found until at least 4–5 years had passed after the FNPP accident. Densities of and reproductive performance by T. clavigera populations near FNPP in 2017 remained below levels before the accident. Although invertebrate larval recruitment from remote areas to the intertidal zone near FNPP could have been expected, this was not clearly observed until 2016 at the earliest. Thus, it is possible that environmental factors inhibited invertebrate reproduction, recruitment or both in the intertidal zone near FNPP at least for 5 years.

Keywords  Fukushima Daiichi Nuclear Power Plant · Intertidal biota · Species richness · Population decline · Population density · Recovery · Recruitment · Reproduction · Rock shell (Thais clavigera) · Sessile organisms
6.1 Introduction

After the 2011 Tohoku earthquake (Mw 9.0) and tsunami on 11 March 2011 (the Great East Japan Earthquake Disaster), three nuclear reactors melted down at Fukushima Daiichi Nuclear Power Plant (FNPP). Hydrogen explosions in reactor buildings resulted in the emission of hundreds of petabecquerels (PBq) of radionuclides to the environment [1]. The amount of radionuclide leakage from the FNPP accident was about one-tenth of that released by the 1986 Chernobyl NPP accident in Ukraine, where the total release of radionuclides was estimated to be 5,300 PBq, excluding radioactive noble gases (e.g., krypton-85, xenon-137 etc) [1].

The Tokyo Electric Power Company (FNPP owner and operator, TEPCO) estimated that 500 PBq of radioactive noble gases, 500 PBq of iodine-131 (131I), 10 PBq of cesium-134 (134Cs) and 10 PBq of 137Cs were released from FNPP to the atmosphere between March 12 and 31, 2011 [2] and that atmospheric fallout and direct leakage from the reactors released an additional 11 PBq of 131I, 3.5 PBq of 134Cs and 3.6 PBq of 137Cs into the marine environment between March 26 and September 30, 2011 [3]. Meanwhile, total deposition of 137Cs from the atmosphere onto the ocean surface is estimated to have been 5–15 PBq [4–8]. Estimates of direct 137Cs leakage from FNPP into the sea range from 3 to 6 PBq [5–13].

Four major sources released FNPP-derived radionuclides to the environment. In the order of decreasing magnitude, these were as follows:

1. The initial venting and explosive releases of gases and volatile radionuclides to the atmosphere, the largest and earliest source, which led to fallout on land and at sea.
2. Direct leakage of contaminated material, including radionuclides from the reactors to the sea during emergency cooling efforts at FNPP.
3. Ongoing radionuclide release via groundwater discharge.
4. Ongoing radionuclide release via river runoff.

Radionuclide releases via groundwater and river discharge were much smaller than the initial atmospheric fallout and the subsequent direct leakage [14].

Although transport models indicate that more than 80% of the atmospheric fallout derived from the FNPP accident would have been on the ocean surface, with the highest levels of deposition expected in coastal waters near FNPP. However, there are no atmospheric fallout data over the ocean to confirm this [14].

The severity of the FNPP accident raised concerns about contamination of aquatic organisms by radionuclides, in both freshwater and marine environments. By the end of March 2011, the Japanese government began to determine activity concentration of radionuclides (i.e., γ-emitters) in aquatic organisms (i.e., fish and shellfish) to protect human consumers. The analytical results are available on the website of the Fisheries Agency of Japan [15]. In general, contamination of marine organisms by radioactive Cs is higher in demersal fish than in pelagic fish; radionuclide contamination levels in both demersal and pelagic fish collected off Fukushima Prefecture have been higher than those in similar fish collected off
other prefectures [15]. The activity concentrations of radioactive Cs in fish tissue, however, have since decreased in most fish sampled from the region (e.g., Wada et al. [16]). Fishing operations in the Fukushima Prefecture region restarted on a trial basis in June 2012, and the areas and species targeted have gradually expanded since then [17].

Since the FNPP accident occurred, there have been many researches and activities in the marine environment, including coastal waters off Fukushima Prefecture, to analyze spatiotemporal changes in radionuclide activity concentration, to determine the final fate of the radionuclides introduced into seawater and to evaluate contamination levels in marine organisms. However, within 20 km of FNPP, there are few available data on the distribution and spatiotemporal changes of radionuclides emitted from FNPP and their possible ecological effects.

According to the literature, wildlife including invertebrates, is tolerant to varying degrees of γ-radiation; at 100–1000 mGy/day, some mortality can be expected in larvae and hatchlings of flatfish [18]. At 10–100 mGy/day, reduced reproductive success is observed in flatfish, and at 1–10 mGy/day, reduced reproductive success due to reduced fertility is possible [18]. Invertebrates such as crabs are more tolerant of radiation than are flatfish [18]. Estimated acute lethal dose (LD₃₀) is >100 Gy for marine invertebrates, 10–25 Gy for fishes, and 0.16 Gy for fish (salmonid) embryos [19]. Chronic exposure has yielded no-observed-adverse-effect dose rate of 10–30 mGy/h (= 240–720 mGy/day) for mortality and 3.2–17 mGy/h (= 76.8–408 mGy/day) for reproduction in snails, marine scallops, clams and crabs [19]. The no-observed-effect dose rate for fish reproduction is 1 mGy/h (= 24 mGy/day) [19]. The United Nations Scientific Committee on the Effects of Atomic Radiation (UNSCEAR) extensively analyzed the relevant data on the effect of radiation on the environment and on nonhuman biota [20, 21], concluding that maximum dose rate of less than 400 μGy/h (= 9.6 mGy/day) to any individual aquatic organism would be unlikely to have any detrimental effects at the population level [22]. This is based on the knowledge that there is little consistent and substantial evidence for any effect on reproduction at dose rate of <200 μGy/h (= 4.8 mGy/day) [23–25]. A generic dose rate of 10 μGy/h (= 240 μGy/day) is suggested for use in screening out environmental exposure situations of negligible concern [25–27].

On the other hand, anaseismic effects and the effect of tsunami backwash on the coastal organisms of eastern Japan have been estimated to be large or serious [28, 29]. The tsunami also caused the loss of tidal flats, as well as subsidence, temporary land elevation, and changes in sediment composition such as sludging or muddying, all of which had either direct or indirect impacts on the benthic organisms inhabiting the coastal areas [30]. The impacts of the tsunami on coastal benthic organisms, and the severity of these impacts, vary among species and sites [28, 29]. Temporary population declines have been observed in the Manila clam (Ruditapes philippinarum) in Matsukawaura, Fukushima Prefecture, although no effects on distribution have been observed in the bivalve Gomphina melanegis in Otsuchi Bay, Iwate Prefecture [31, 32]. The tsunami also affected reef resources, including abalone and sea urchins, through loss or reduction of seaweed forests as well as through direct
impacts, though in Miyagi and Iwate Prefectures, the populations of some of these species have been observed to recover [32, 33].

Although dose rate from the environmental exposure in intertidal invertebrates near FNPP after the accident has not yet been estimated, Horiguchi et al. [34] investigated the intertidal zone from Shirahama (Chiba Prefecture) to Kuji (Iwate Prefecture) along the coastline of eastern Japan (a total distance of ca. 800 km) in 2011, 2012 and 2013 to evaluate the ecological effect of the FNPP accident that accompanied the Great East Japan Earthquake Disaster. They observed that the number of intertidal species decreased significantly with proximity to FNPP and that no rock shell (Thais clavigera) specimens were collected near FNPP, from Hirono to Futaba Beach (a distance of approximately 30 km) in 2012. The collection of rock shell specimens at many other sites hit by the tsunami suggested that the absence of rock shells around FNPP in 2012 might have been caused by the FNPP accident in 2011. Quantitative surveys in 2013 showed that the number of species and population densities in the intertidal zones were much lower at sites near or within several kilometers south of FNPP than at other sites and lower than in 1995, especially in the case of Arthropoda. There is no clear explanation for these findings, but evidently, the intertidal biota around FNPP has been harmed since the nuclear accident [34].

In June 2014, May and June 2015, and June 2016, we conducted quantitative quadrat surveys of sessile invertebrates at seven intertidal sites previously surveyed in Ibaraki, Fukushima and Miyagi Prefectures including sites near FNPP, to evaluate whether the intertidal invertebrates had recovered from the declines that followed the accident; we measured population densities (i.e., abundance, as indicated by the number of individuals per square meter), biomass (total wet weight of invertebrates per square meter) and species numbers. Additionally, in April, July, August, and September from 2012 to 2017, we monitored the density of rock shell (T. clavigera) populations and counted their egg capsules in the field near FNPP.

6.2 Materials and Methods

On June 13-16 and 26-28, 2014, May 18-21 and June 17-19, 2015, June 4-7, 21, 23 and 24, 2016, quantitative quadrat surveys of sessile invertebrates were conducted at seven sites: Tomioka Fishing Port (Tomioka Town; shown as Tomioka in Figs. 6.1b, 6.2, 6.3, 6.4, 6.5, 6.6 and 6.7), Ottozawa coast (Okuma Town; shown as Okuma in Figs. 6.1b, 6.2, 6.3, 6.4, 6.5, 6.6 and 6.7), Kubo-yaji coast (Futaba Town; shown as Kuboyaji in Figs. 6.1b, 6.2, 6.3, 6.4, 6.5, 6.6 and 6.7) and Urajiri coast (Minami-Soma City; shown as Urajiri in Figs. 6.1b, 6.2, 6.3, 6.4, 6.5, 6.6 and 6.7) in Fukushima Prefecture within a 20-km radius of FNPP; Hasaki Beach (Kamisu City) and Kujihama Beach (Hitachi City) in Ibaraki Prefecture; and Watanoha coast (Ishinomaki City; shown as Ishinomaki in Figs. 6.1b, 6.2, 6.3, 6.4, 6.5, 6.6 and 6.7) in Miyagi Prefecture as reference sites for comparison (Fig. 6.1b and Table 6.1). These sites were the same as those surveyed in 2013 and were representative of
those used in the 2012 surveys in terms of substrate (i.e., tetrapods or similar concrete structures set along the coast for wave protection) as well as distance from FNPP [34]. Sessile organisms on the surface of tetrapods or similar concrete structures within a 50 × 50-cm quadrat were collected at three different elevations in the intertidal zone (the lower, middle and upper intertidal zones) at each site. Specimens were preserved in neutral buffered 10% formalin solution. After identifying the species, the number of individuals and wet weight were determined for each species on the spot, along with the sampling location and elevation. Additionally, we conducted field observations on the density and spawning behavior of *Thais clavigera* populations near FNPP in April, July, August and September from 2012 to 2017 (Fig. 6.1c and Table 6.2). The density of *T. clavigera* populations was expressed by the number of individuals collected per minute, because it was difficult to quantitatively represent
Fig. 6.2 Total numbers of sessile invertebrate species in the intertidal zone of seven sites from 2013 to 2016, as sampled with a 50 × 50-cm quadrat. (a) 2013, (b) 2014, (c) 2015 and (d) 2016. FNPP is located between Okuma and Kubo-yaji. Tomioka, Okuma, Kubo-yaji and Urajiri are located within the 20-km radius of FNPP. Distances between sites do not correspond to bar locations on the charts.

Fig. 6.3 Densities of sessile invertebrates (number/m²) in the intertidal zone (by elevation) from 2013 to 2016. (a) 2013, (b) 2014, (c) 2015 and (d) 2016. L lower intertidal zone, M middle intertidal zone, U upper intertidal zone. Data were collected with a 50 × 50-cm quadrat. FNPP is located between Okuma and Kubo-yaji. Tomioka, Okuma, Kubo-yaji and Urajiri are located within the 20-km radius of FNPP. Distances between sites do not correspond to bar locations on the charts. Pink dotted line represents the average number of individuals/m² from quadrat surveys of sessile organisms conducted in May 1995 at 20 sites along the coast of Fukushima Prefecture [37]: the average population density in 1995 was 7,158 individuals/m², consisting of Arthropoda (4593, 64.2%), Annelida (179, 2.5%), Mollusca (2348, 32.8%) and other organisms (38, 0.5%).
To assess similarity in the biotic community structure of the intertidal zone in the survey from 2014 to 2016, we conducted cluster analysis (group average method) on Bray-Curtis similarity matrices for the number of species, population density (the number of individuals/m²) and total sessile biomass (whole wet weight per square meter, g/m²). Population density data from the lower, middle, and upper intertidal zones were merged within each site, and log (1 + x) transformed for calculation of Bray-Curtis similarity. Site grouping was performed with a cut-off similarity level of 70 or 80%. Differences in the biotic community structure represented by Bray-Curtis similarity among site groups [35] were tested by analysis of similarity (ANOSIM). Possible differences in population densities of Arthropoda (i.e., the number of individuals/m²) at sites near FNPP (i.e., Tomioka, Okuma, Kuboyaji) compared to those at other sites (i.e., Hasaki, Kujihama, Urajiri, and Ishinomaki) were evaluated using a t-test under the assumption that they had approximately equal variances each year. Statistical analyses (t-test) were performed with Microsoft Excel 2016 statistical software, except for cluster analysis and ANOSIM, which were conducted with PRIMER6 software [36].
Fig. 6.5 Dendrograms constructed by the group average method on Bray-Curtis similarity matrices for (a) number of species, (b) population density, and (c) total biomass of intertidal invertebrates collected in northeastern Japan on June 13–16 and 26–28, 2014. Population density data from the lower, middle, and upper intertidal zones were merged within each site, and log (1+x) transformed for calculation of Bray-Curtis similarity. Site groupings based on cut-off similarity level of 70% (dotted lines) are shown as S1-S2, N1-N2, and W1-W3 for panels (a), (b), and (c), respectively. The two sites (Okuma and Tomioka) located south of FNPP are shown in red text.
Fig. 6.6 Dendrograms constructed by the group average method on Bray-Curtis similarity matrices for (a) number of species, (b) population density, and (c) total biomass of intertidal invertebrates collected in northeastern Japan on May 18–21 and June 17–19, 2015. Population density data from lower, middle, and upper intertidal zones were merged within each site, and log (1+x) transformed for calculation of Bray-Curtis similarity. Site groupings based on cut-off similarity level of 80% (dotted lines) are shown as S1-S2, N1-N2, and W1-W3 for panels (a), (b), and (c), respectively. The two sites (Okuma and Tomioka) located south of FNPP are shown in red text.
Fig. 6.7 Dendrograms constructed by the group average method on Bray-Curtis similarity matrices for (a) number of species, (b) population density, and (c) total biomass of intertidal invertebrates collected in northeastern Japan on June 4–7, 21, 23 and 24, 2016. Population density data from the lower, middle, and upper intertidal zones were merged within each site, and log (1+x) transformed for calculation of Bray-Curtis similarity. Site groupings based on cut-off similarity level of 80% (dotted lines) are shown as S1-S2, N1-N4, and W1-W4 for panels (a), (b), and (c), respectively.
6.3 Results

In the intertidal zone at all sites surveyed from 2014 to 2016, the sessile species composition was dominated by the Mollusca and Arthropoda, similar to the 2013 results [34] (Fig. 6.2). The maximum number of intertidal species in 2014 was 35 at Hasaki, followed by 29 at Urajiri. The minimum of five species was at Okuma, located approximately 1 km south of FNPP (Fig. 6.2b). However, the similarity in species number between Tomioka and Okuma, both of which are located south of FNPP, was not very high and did not significantly differ from that between other sites in 2014, as determined by analysis of similarity (ANOSIM) (\( P = 0.14 \); Figs. 6.2b and 6.5a). This result differed from that of the 2013 survey. The pattern of more species being identified in 2014 at Hasaki, Kujihama, Kubo-yaji, Urajiri and Ishinomaki than at Tomioka and Okuma was also observed in 2015 (Fig. 6.2c). The similarity in species number between Tomioka and Okuma, south of FNPP, was high and differed significantly from that between other sites in 2015 (\( P < 0.05 \); Figs. 6.2c and 6.6a). In 2016, however, no such pattern of more species being identified at Hasaki, Kujihama, Kubo-yaji, Urajiri and Ishinomaki than at Tomioka and Okuma was observed, due to an increase in the number of species at Tomioka and Okuma (Fig. 6.2d). The similarity in species number between Tomioka and Okuma, located south of FNPP, was not very high and differed insignificantly from that between other sites in 2016 (\( P = 0.29 \); Figs. 6.2d and 6.7a).

Population densities were higher in the lower and middle intertidal zones than in the upper intertidal zone at all sites surveyed from 2014 to 2016 (Fig. 6.3b–d), which is similar to the 2013 results [34]. Mollusca (e.g., mussels such as *Mytilus galloprovincialis* and *Septifer virgatus*) and Arthropoda (e.g., barnacles such as...
Table 6.2  Site locations for surveys on density and spawning behavior of *Thais clavigera* populations from 2012 to 2017

| Date       | 2017 Survey | 2015 Survey | 2014 Survey | 2013 Survey | 2012 Survey | Sampling Location | Town/City | Name of site | Latitude | Longitude |
|------------|-------------|-------------|-------------|-------------|-------------|-------------------|-----------|--------------|----------|-----------|
| Naraha     | 26-Apr      | 8-Apr       | 19-Apr      | 15-Apr      | 24-Apr      | Naraha           | Yamadahama | 37°15'29.1″N | 141°00'50.0″E |
| Naraha     | 26-Apr      | 8-Apr       | 19-Apr      | 15-Apr      | 24-Apr      | Naraha           | Shimo-shigeoka | 37°16'39.1″N | 141°01'04.6″E |
| Naraha     | 26-Apr      | 8-Apr       | 19-Apr      | 15-Apr      | 24-Apr      | Naraha           | Namikura   | 37°18'25.2″N | 141°01'28.1″E |
| Tomioka    | 27-Apr      | 11-Apr, 19-Jul | 21-Apr, 30-Jul, 31-Aug | 16-Apr, 13-Jul, 11-Aug | 9-Aug       | Tomioka          | Kegaya     | 37°19'40.5″N | 141°01'35.1″E |
| Tomioka    | 27-Apr, 25-Jul, 22-Aug | 11-Apr, 19-Jul, 20-Aug | 21-Apr, 30-Jul, 31-Aug | 16-Apr, 13-Jul, 11-Aug, 10-Sep | 9-Aug       | Tomioka          | Tomioka Fishing Port | 37°20'16.3″N | 141°01'45.3″E |
| Okuma      | 28-Apr, 22-Jul | 9-Apr, 20-Jul, 18-Aug | 20-Apr, 31-Jul | 13-Jul, 11-Aug, 10-Sep | 9-Aug       | Okuma            | Kumagawa   | 37°23'16.7″N | 141°02'03.0″E |
| Okuma      | 28-Apr, 25-Jul, 22-Aug | 9-Apr, 20-Jul, 18-Aug | 20-Apr, 31-Jul, 31-Aug | 17-Apr, 13-Jul, 11-Aug, 10-Sep | 9-Aug       | Okuma            | Koirino    | 37°23'26.5″N | 141°02'04.5″E |
| Okuma      | 29-Apr, 26-Jul, 23-Aug | 10-Apr, 21-Jul, 19-Aug | 22-Apr, 1-Aug, 1-Sep | 17-Apr, 14-Jul, 12-Aug, 11-Sep | 10-Aug      | Okuma            | Ottozawa   | 37°24'21.2″N | 141°02'00.4″E |
| Futaba     | 29-Apr, 26-Jul, 23-Aug | 10-Apr, 21-Jul, 19-Aug | 22-Apr, 1-Aug, 1-Sep | 17-Apr, 14-Jul, 12-Aug, 11-Sep | 10-Aug      | Futaba           | Kuboyaji   | 37°26'41.4″N | 141°02'10.9″E |
| Date       | Dates       | Sampling Site                  | Location            | Latitude      | Longitude     |
|------------|-------------|--------------------------------|---------------------|---------------|---------------|
| 29-Apr     | 23-Aug      | Futaba beach                  | 37°27′12.1″N 141°02′20.7″E |
| 29-Apr     |             | 18-Apr, 14-Jul                 | Ukedo fishing port  | 37°28′51.7″N 141°02′35.6″E |
| 30-Apr     | 26-Jul      | Minami-Soma Urajiri           | 37°31′33.7″N 141°01′53.8″E |
| 30-Apr     | 12-Apr      | Minami-Soma Tsunobeuchi       | 37°32′39.3″N 141°01′43.2″E |
|            | 19-Apr      | Minami-Soma Murakami          | 37°34′22.4″N 141°01′34.4″E |

FNPP is located between Ottozawa and Kuboyaji. All sites were located within a 20-km radius of FNPP.
*Chthamalus challengeri* predominated at almost all sites surveyed. Species composition in 2014 was highly similar between Tomioka and Okuma, both located south of FNPP ($P < 0.05$; Figs. 6.3b and 6.5b), but not in 2015 ($P = 0.14$; Figs. 6.3c and 6.6b). In 2016, the species composition at Kujihama, Urajiri and Kubo-yaji differed significantly from that at the other sites (Tomioka, Okuma, Hasaki and Ishinomaki) ($P < 0.05$; Figs. 6.3d and 6.7b), but not from those at sites south of FNPP (i.e., Tomioka and Okuma), although it was one of typical characteristics in the 2013 survey that population densities in the intertidal zone were significantly lower at sites south of FNPP than at other sites [34].

From 2014 to 2016, the population density of sessile organisms at Kubo-yaji, approximately 1 km north of FNPP, was greater than those at Kujihama and Urajiri (Fig. 6.3b–d). However, the species composition of sessile organisms at Kubo-yaji differed markedly from those at other sites surveyed in that the Arthropoda accounted for less than 1% of all sessile organisms collected (Fig. 6.3b–d), which is similar to the 2013 survey [34]. Population densities of Arthropoda at sites near FNPP (Tomioka, Okuma, and Kubo-yaji) were significantly lower than those at other sites in 2014 and 2015 ($t$-test; $P < 0.05$), but not in 2016 ($P = 0.66$).

The combined wet weight of all sessile organisms from 2014 to 2016 was the greatest in the lower intertidal zone at almost all sites, followed by the middle and upper intertidal zone (Fig. 6.4b–d), similar to the results in 2013 [34]. From 2014 to 2016, molluscan biomass predominated at all surveyed sites, followed by arthropodan biomass (Fig. 6.4b–d). Species composition at Tomioka was similar to that at Okuma and differed significantly from that at other sites in 2014 and 2015 ($P < 0.05$; Figs. 6.4b, c, 6.5c, and 6.6c). In 2016, however, there was a significant difference of similarity in the species composition between the occurrences at Kubo-yaji, Kujihama, most central Fukushima sites (Okuma, Tomioka, and Urajiri), and the sites at both ends of the survey (Hasaki and Ishinomaki) ($P < 0.05$; Figs. 6.4d and 6.7c), but not in the south of FNPP (i.e., Tomioka and Okuma), which was one of the typical characteristics observed in the 2013 survey: total sessile biomass in the intertidal zone were significantly lower at sites south of FNPP than at other sites in the 2013 survey [34].

The distribution of rock shells (*T. clavigera*) near FNPP gradually expanded from sites in north (Minami-Soma City and Namie Town Fig. 6.8a, b, respectively) and south (Naraha Town and Tomioka Town Fig. 6.8f, e, respectively) to sites in central Fukushima (Futaba Town and Okuma Town Fig. 6.8c, d, respectively). Densities also seemed to increase from 2012 to 2016, with large variations (Fig. 6.8). When we started the field survey in December 2011, 9 months after the FNPP accident, *T. clavigera* had disappeared from Ottozawa, Okuma Town, approximately 1 km south of FNPP; they remained absent until July 2016: the first record on the density of *T. clavigera* population at Ottozawa, Okuma Town, in July 2016 was 0.03 individuals collected per minute (i.e., 3 individuals collected within 90 min). Consequently, the gap in the distribution of *T. clavigera* along the Fukushima coast seems to have closed by July 2016. However, population densities have remained low (around or less than 1 individual collected per minute) at
sites in Okuma Town, within a few kilometers south of FNPP (Fig. 6.8d). Moreover, neither spawning behavior nor egg capsules were observed in populations at sites near FNPP (from Tomioka Fishing Port to Kubo-yaji) through the summer of 2016; egg capsules were first observed at Koirino, Ottozawa (Okuma Town) and Kubo-yaji (Futaba Town) in the summer of 2017 (Table 6.3 and Fig. 6.1c).
6.4 Discussion

In 2011, 2012 and 2013, Horiguchi et al. [34] investigated the ecological effect in the intertidal zone of eastern Japan affected by the FNPP accident that accompanied the Great East Japan Earthquake Disaster. They reported that the number of intertidal species decreased significantly with proximity to FNPP, and that no rock shell (T. clavigera) specimens were found near FNPP, from Hirono to Futaba Beach (a distance of approximately 30 km) in 2012. The presence of rock shell specimens at many other sites hit by the tsunami suggested that the absence of rock shells around FNPP in 2012 might have been caused by the FNPP accident. The quantitative 2013 surveys also showed that the number of species and population densities in the intertidal zones were much lower at sites near or within several kilometers south of FNPP than at other sites, and lower than in 1995, especially in the case of Arthropoda. There is no clear explanation for these findings. Although there are no estimates of radiation doses from environmental exposure in invertebrates in the intertidal zone near FNPP after the accident, clearly the intertidal biota around FNPP has been harmed since the nuclear accident.

The present study was carried out to investigate temporal changes in species composition and population densities in the intertidal zone of eastern Japan since 2014. Increases in species richness were evident at Tomioka Fishing Port (Tomioka) and Ottozawa (Okuma), which are located south of FNPP, from 2016, though this increase might have started at Ottozawa (Okuma) in 2015 (Fig. 6.2). Meanwhile, population densities have increased in the lower and middle intertidal zones at Ottozawa (Okuma) and Tomioka Fishing Port (Tomioka) since 2015 and 2016, respectively (Fig. 6.3). These increases were due to increases in small mussels (i.e., Mytilus galloprovincialis) at Ottozawa (Okuma) and barnacles (i.e., Semibalanus cariosus and Chthamalus challengeri) at Tomioka Fishing Port (Tomioka). In 1995, TEPCO conducted similar seasonal surveys using 30 × 30-cm quadrats at 20 sites in intertidal zone along the coast of Fukushima Prefecture, but only published a summary of their results [37]. In May 1995, they found an average of 7,158 individual sessile organisms/m², consisting of Arthropoda (4,593, 64.2%), Annelida (179, 2.5%), Mollusca (2348, 32.8%), and other organisms (38, 0.5%) [37]; thus, although Arthropoda clearly predominated, many other invertebrates were also present in Fukushima Prefecture in 1995, before the FNPP accident. Compared with the TEPCO data, we observed similar or higher numbers of individuals/m² in the lower and middle intertidal zones at Ottozawa (Okuma) in 2015 and Tomioka Fishing Port (Tomioka) in 2016 (Fig. 6.3). Because of their small sizes, however, increases in the total biomass at Ottozawa (Okuma) in 2015 and Tomioka Fishing Port (Tomioka) in 2016 were unclear (Fig. 6.4).

The species composition of sessile organisms at Kubo-yaji, approximately 1 km north of FNPP, differed markedly from that at other sites surveyed, in that the Arthropoda accounted for less than 1% of all sessile organisms collected (Fig. 6.3b–d). Although Arthropoda predominated in the intertidal zone of Fukushima Prefecture in 1995 [37], population densities seem to have decreased
after the FNPP accident. The reduced population densities there have continued since 2013. Population densities of Arthropoda at sites near FNPP (Tomioka, Okuma, and Kubo-yaji) were significantly lower than those at other sites in 2014 and 2015 (t-test; \( P < 0.05 \)), but not in 2016 (\( P = 0.66 \)), possibly due to increases in barnacle population densities in the lower intertidal zone at Tomioka Fishing Port (Tomioka) in 2016 (Fig. 6.3d).

No \( T. \ clavigera \) specimens were observed near FNPP from Hirono to Futaba Beach (a distance of approximately 30 km) in 2012 [34]. In the present study conducted during 2014–2017, however, their distribution gradually expanded year by year from sites in the north and south to sites along the central Fukushima coast (Fig. 6.8), and the gap in the distributions, where no \( T. \ clavigera \) specimens were found, had closed by July 2016, approximately a year after an observation of increases in small mussels (i.e., \( Mytilus \ galloprovincialis \)) at Ottozawa (Okuma) in 2015; these mussels are prey for \( T. \ clavigera \) (Fig. 6.3c). The increase in prey organisms, such as small mussels (i.e., \( Mytilus \ galloprovincialis \)) at Ottozawa (Okuma) in 2015 might have been associated with the first observation of \( T. \ clavigera \) specimens there in July 2016, since our field survey started in December 2011, 9 months after the FNPP accident. The densities of \( T. \ clavigera \) populations seemed to increase from 2012 to 2016, with large variations (Fig. 6.8). Low densities in \( T. \ clavigera \) populations (around or less than 1 individual collected per minute), however, have persisted at sites in Okuma Town, within a few kilometers south of FNPP (Fig. 6.8d). In general, population densities of \( T. \ clavigera \) in Japan range from 1 to 4 individuals collected per minute (Horiguchi et al., unpublished data). Neither spawning behavior nor egg capsules spawned by \( T. \ clavigera \) populations were observed at sites near FNPP (from Tomioka Fishing Port to Kubo-yaji) through the summer of 2016 (Table 6.3). Egg capsules were first observed at Koirino, Ottozawa (Okuma Town), and Kubo-yaji (Futaba Town) in the summer of 2017, though the number of egg capsules spawned was limited (Table 6.3).

The number of species and population densities in the intertidal zone near FNPP did not begin to increase until at least 4–5 years after the FNPP accident. Even in 2017, 6 years after the disaster, densities of and reproductive performance by \( T. \ clavigera \) populations near FNPP remained below pre-disaster levels. Most invertebrates in the intertidal zone, including sessile organisms, have planktonic stages in their early life histories for certain periods: for example, \( T. \ clavigera \) has a veliger larva that lives a few months before settling [38]. In the coastal waters off Fukushima Prefecture, the generally weak currents and tides [39] are expected to effect the dispersion of larvae of most invertebrates in the intertidal zone, as well as dispersion of FNPP-derived radioactive Cs. The Kuroshio and Oyashio currents, which are the wind-driven boundary currents of the North Pacific, meet in the Kuroshio–Oyashio transition area off FNPP [40]. The Kuroshio is part of the subtropical gyre and transports warm saline water along the south coast of Japan and then eastward via the Kuroshio Extension (KE), whereas the Oyashio is part of the subarctic gyre and transports cold less saline water southward [41]. Actually, results on the bottom-sediment survey collected from a coastal strip (~30 × 120 km) off FNPP in October 2012 revealed that radioactive Cs concentration in the
uppermost sediment layer were higher south of FNPP, though high activity concentration patches were also observed at a few sites north of FNPP [42]. Therefore, recruitment of their larvae could have been expected from remote areas to sites near FNPP. Nevertheless, data in the present study revealed that neither juvenile nor adult *T. clavigera* were observed at Ottozawa (Okuma), approximately 1 km south of FNPP, until at least 2016; thus, for invertebrates including sessile organisms in the intertidal zone near FNPP, some environmental factors seem to have inhibited reproduction, recruitment, or both.

Among the factors that might have inhibited recovery of invertebrates are strong waves; sustainability or suitability of substrate (e.g., tetrapods or similar concrete structures set along the coast for wave protection) for larvae to settle on; harmful substances such as radionuclides; heavy metals; turbidity or suspended material; insufficient amounts of prey organisms; and excess of predators. Before the FNPP accident, there were strong waves, substrate (e.g., tetrapods or similar concrete structures set along the coast for wave protection) for larvae to settle on, and various kinds of intertidal predator and prey invertebrates in the coastal waters off Fukushima Prefecture, similar to those after the FNPP accident. Although the possibility of changes in predators or prey affecting invertebrate populations cannot be ruled out, strong waves and substrate are unlikely to be those that inhibited reproduction or recruitment of invertebrates including sessile organisms in the intertidal zone near FNPP. Regarding the possible impacts by harmful substances, such as radionuclides, heavy metals and turbidity or suspended material, we need to carefully examine the effect of construction to cover the radionuclide-contaminated bottom sediments of the FNPP port; this construction was conducted between March 2012 and April 2015 [43, 44]. Approximately 33,000 m$^3$ of bentonite and cement were used for the construction from March 14 to July 5, 2012 [43]. Another approximately 32,500 m$^3$ of bentonite, sand, and cement were also used for the construction from July 17, 2014 to April 23, 2015 [44]. Cement is known to include several heavy metals, and the specific gravity of cement is 3050 kg/m$^3$ [45]. We tried to estimate the total amount of heavy metals, which had been deposited into the marine environment during the construction. Assuming that the half of total amount of material (i.e., bentonite, sand and cement) used for the construction was cement, then the total weight of cement used for construction to cover the radionuclide-contaminated bottom sediments of the FNPP port was approximately 100,000 metric tons. This cement included various amounts of chromium (Cr), copper (Cu), zinc (Zn) and lead (Pb) and other heavy metals, which were intentionally deposited into the marine environment adjacent to FNPP; consequently, the coastal waters of Fukushima Prefecture, especially close to FNPP, might have been contaminated (Table 6.4). Thus, these heavy metals, as well as turbidity or suspended material from the construction activity, may have had ecological effects (i.e., adverse effects on larval, juvenile, and adult invertebrates in the intertidal zone) that should be elucidated in the near future. Similar impacts on the number of species and population densities by the restoration work of coastal structures were also likely at Tomioka Fishing Port (Tomioka) (since 2015), Urajiri (since 2013) and Ishinomaki (since 2012) [46–48]. The lack of observed egg capsules spawned by rock shells at Urajiri
There have been many scientific researches and review papers on the toxicities of heavy metals to marine and estuarine organisms [49–54]. Although it is known that the toxicities of heavy metals to marine and estuarine organisms can vary according to their chemical speciation (e.g., free ions are more toxic than other chemical species) and pH [50, 53, 54], the embryos and larvae of marine or estuarine invertebrates are generally sensitive to heavy metals: for example, LC50 values (the concentrations at which 50% of the tested organisms die) for the embryos of American oyster (*Crassostrea virginica*) were 5.6 ppb (µg/L), 5.8 ppb (µg/L), and 103 ppb (µg/L) for mercury (Hg), silver (Ag) and Cu, respectively. LC50 values for nickel (Ni), Pb and cadmium (Cd) were 1.18 ppm (mg/L), 2.45 ppm (mg/L), and 3.80 ppm (mg/L), respectively [52]. Moreover, in the presence of multiple metals, the toxicities are reported to be additive; the presence of multiple metals may be more realistic than the presence of single metals in marine/estuarine environments [54].

The effect of turbidity and suspended material in aquatic environments has been well studied (e.g., Stern and Stickle [55]). Most studies on adult estuarine and marine bivalves (clams, oysters, and mussels) have indicated that the mortality rate among populations adjacent to dredging and disposal areas is low, except for individuals directly buried by the disposal operation; however, the percentage occurrence of normally developing eggs and larvae may decrease as the concentration of suspended solids increases in the range of concentrations normally resulting from dredging and disposal [55]. Thus, further experimental studies are necessary to elucidate the effect of construction to cover the radionuclide-contaminated bottom sediments of the FNPP port, as well as other restoration work of coastal structures.

The exposure situation for intertidal organisms in Fukushima Prefecture could be complex with many aspects, including various potential direct impacts such as physical harm from the tsunami, and toxicity from chemicals and radionuclides in

| Metals      | Composition (mg/kg) | Estimated total amount (t) |
|-------------|---------------------|---------------------------|
| Total Cr    | 98                  | 9.8                       |
| Hexavalent Cu | 10.8                | 1.1                       |
| Cu          | 140                 | 14.0                      |
| Zn          | 511                 | 51.0                      |
| As          | 18.9                | 1.9                       |
| Se          | <1                  | <0.1                      |
| Cd          | 2.0                 | 0.2                       |
| Total Hg    | 0.023               | 0.002                     |
| Pb          | 111                 | 11.1                      |

General composition of metals as reported by [45]. We estimated the total amount of cement used for the construction to be approximately 100,000 metric tons, assuming that cement comprised half of the total amount of material (i.e., bentonite, sand and cement) used

in 2015 (Table 6.3) might also have been associated with the restoration of coastal structures.

Table 6.4 General composition of heavy metals in cement and their estimated total amounts used in the construction to cover the radionuclide-contaminated bottom sediments of the FNPP port from March 2012 to April 2015
the massive release immediately after the accident, potentially leading to acute effects [34]. Thereafter, there could have been continued releases to the sea of radio-
uclides and other harmful substances (i.e., heavy metals and turbidity or suspended material from the construction to cover the radionuclide-contaminated bottom sedi-
ments of the FNPP port); these releases could have had ecological effects, for exam-
ple, on interspecific relationships (such as predator-prey relationships or competition
for prey organisms and habitat). The effect could also have involved intraspecific
relationships (competition for prey organisms, habitat and mating partners) [56–58].

At much higher dose rates, possibly immediately after the accident, differences
among taxa in sensitivity to radiation could create competitive advantages for resis-
tant organisms within a taxon and between populations of interacting taxa [23, 59].
Thus, in addition to differences in the direct radiosensitivity of individual organ-
isms, life history traits including responses to changes in resources and generation
times affect the consequences of radiation exposure. Exposures to radiation may
have been high during or shortly after the accident [56]. In addition, note that par-
ticularly at lower doses, ecological factors and variability can be more important
than direct radiation effects; therefore, a different conceptual methodology may be
necessary to assess ecosystem-level effects, possibly including site-specific assess-
ment of the effect of potential disturbances on ecosystems [56].

Further studies are needed to clarify the main causal factors for declining inter-
tidal biota and subsequent slow recovery near FNPP, possibly through determining
the acute, subacute and chronic toxicity of various radionuclides, chemicals and
other factors in laboratory experiments. Continued field observations of spatiotem-
poral changes in the populations of sessile organisms near FNPP, including rock
shell populations, are also necessary to evaluate their recovery in the future; these
studies should consider the characteristics of habitats that may influence the distri-
bution of sessile organisms. The focus should be on increasing population densities
and reproductive success in terms of active behaviors such as mating and egg-laying
and the subsequent successful recruitment of larvae and juveniles. Both field and
laboratory studies will also be necessary to observe and evaluate possible multigen-
erational effects such as changes in reproductive success resulting from exposure to
low-dose radiation and other environmental stressors.

6.5 Conclusions

We conclude that the population densities and species richness of intertidal inverte-
brates along the coast off Fukushima Prefecture have decreased since March 2011,
especially south of FNPP. The recovery from this decline in population densities of
intertidal invertebrates, including T. clavigera, as well as in the number of inverte-
brate species, has been limited and slow at sites located near and south of FNPP;
进一步研究需要澄清导致潮间带生物群落减少的主要原因，以及随后的缓慢恢复
特别是在福岛第一核电站附近的地区。这些研究应该考虑栖息地的特性，这可能影响潮
间带生物的分布。研究的焦点应集中在增加潮间带生物群落密度和繁殖成功率，在内
容行为如交配和产卵以及随后成功招募的幼虫和幼鱼。野外和实验室研究也将必要
于观察和评估可能的多代效应，如低剂量辐射暴露和其它环境压力的结果带来的
低剂量辐射暴露和其它环境压力。
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