Oxidative injury caused by individual and combined exposure of neonicotinoid, organophosphate and herbicide in zebrafish

Saurabh Shukla a, Reena C. Jhamtani a, M.S. Dahiya b, Rakhi Agarwal a,⁎

a Laboratory of Analytical & Molecular Toxicology (Forensic Chemistry & Toxicology Laboratory), Institute of Forensic Science, Gujarat Forensic Sciences University, Sector 09, Gandhinagar, 382007, Gujarat, India
b Institute of Forensic Science, Gujarat Forensic Sciences University, Sector 09, Gandhinagar, 382007, Gujarat, India

ARTICLE INFO

Keywords:
Imidacloprid
Dichlorvos
Atrazine
Oxidative stress
Environmental toxicology

ABSTRACT

The greatest challenge in environmental toxicology is to understand the effects of mixture toxicity as environmental pollutants co-exist and exhibit combined effects. Thus, it is necessary to evaluate the mixture toxicity associated with two or more co-existing compounds. Pesticides are widely used to control pest, they are ubiquitous in nature and present in all environmental components. Pesticide residue can be detected in almost all components of environment and food samples. Imidacloprid (IMD) (neonicotinoid), dichlorvos (DIC) (organophosphates) and atrazine (ATZ) are three widely used pesticides for commercial uses. Present work includes the assessment of effects of individual exposure of IMD (27.5 mg/L), DIC (15 mg/L), and ATZ (03 mg/L) and in combination of three (CMD) (13.75 + 7.5 + 1.5 mg/L IMD, DIC & ATZ, respectively) in terms of LPO, GSH content and antioxidant enzymes activities (superoxide dismutase, catalase and glutathione peroxidase) in zebrafish (Danio rerio), exposed for 24 h. CMD group exhibits highest lipid peroxidation than other individually exposed groups. Similarly, the activities of antioxidant enzymes were highest in CMD group with reduced GSH content. Results indicate that exposure to mixture of pesticides develops synergistic effects which were more toxic in comparison to individual exposure and also produce toxicity in all examined tissues rather than selective organ toxicity.

1. Introduction

In natural environment, human being and other living organisms get exposed to the mixture of pesticides, unlike single pesticide. Pesticides are widely used to increase agricultural yield and consequently, raised environmental concern and frequently detected in water bodies, soil, food and vegetables [1]. Pesticides are known to induce reactive oxygen species (ROS) and cause oxidative stress leading to inactivation of antioxidant enzymes and reduction of free radical scavengers. Superoxide dismutase (SOD), catalase (CAT) and glutathione peroxidase (Gpx) are important endogenous antioxidant enzymes that protect cell from oxidative damage [2]. Zebrafish, an important model organism having high predictivity with human involved in various research activities including toxicological research [3,4].

Imidacloprid (IMD), a neonicotinoid pesticide widely used for protecting crops from insects and comparatively safer, thus, considered as a replacement of organophosphate pesticide. It affects insect’s postsynaptic nicotinic acetylcholine receptor and act as an agonist resulting in death [5]. A dose of 10 mg/kg/day for 90 days is not sufficient to produce significant biochemical alterations in experimentally exposed rats [6]. However, research has raised concern over the use of imidacloprid as it affects egg-shell thinning, reduction in egg production and hatching time in birds [7]. Dichlorvos (DIC) is an organophosphate pesticide, widely used for agriculture purpose, household and veterinary uses. Significant concern was raised over the use of DIC due to its acute and chronic toxicity because it is stable in water, soil and air, with a half-life ranging from two (air) to seven days (water and soil) [8]. DIC is extremely toxic to aquatic animals especially fish and known to cause oxidative stress, injury to cell, lipid peroxidation and other events consistent with elevated ROS levels [9]. Atrazine (ATZ) is an s-triazine, commonly used herbicide that inhibits photosynthesis in broadleaf and grass on crops. It is commonly applied to corn, sugarcane, sorghum and turf grass. The widespread use of atrazine has resulted in the contamination of surface and ground water. Atrazine is resistant to degradation and acts as an endocrine disruptor in zebrafish and other animal. Atrazine is also known to cause oxidative stress, injury to cell, lipid peroxidation and other events consistent with
elevated ROS levels [6,10,16].

In mixture toxicity, toxicokinetic and toxicodynamic exchange can occur among toxicants resulting in potentiating and addition effects [11]. Thus, the major concern for this work is to understand the combined effects of mixture toxicity caused by above mentioned pesticides because of their high use and detected in river water samples. Reported minimum concentration of imidacloprid in Ebro river, Spain is 1.64 ng/L and maximum concentration is 14.96 ng/L. Similarly the minimum concentration of atrazine is 8.13 ng/L and maximum is 12.22 ng/L. The study provides huge data on different organophosphate pesticides concentration, however the results of dichlorvos are not reported [12]. Ample studies were conducted on chronic toxicity with these pesticides [6,10,15–17], however, literature is not available on the acute toxicity caused by said pesticides. Therefore, investigations were carried out for the evaluation of the individual and combined effects of acute exposure of imidacloprid, dichlorvos and atrazine in terms of biochemical alterations in zebrafish.

2. Materials and methods

2.1. Chemicals and reagents

Commercial product of Imidacloprid (PubChem CID: 86418), Dichlorvos (PubChem CID: 3039) and Atrazine (PubChem CID: 2256) were procured from pesticide testing laboratory, Gandhinagar, India. All standards and other reagents used were of analytical grade and handled in accordance with good animal practice as defined by the animal welfare bodies and the university committee No. PhD/FS/RA/02. Zebrafish were kept under controlled condition with temperature (25 ± 1 °C), pH 7.0 ± 0.2, conductance alkalinity 0.25 μS cm−1 and 12 h light and dark cycles. Animals were not fed 24 h prior to experimentation. The rationale for pesticides concentration was based on pilot experiments and literature. For, imidacloprid, no data is available for zebrafish. However, the dose of 10 mg/kg/day for rats was observed as no observed effect level (NOEL) [6]. Thus, an experiment was conducted and no mortality was observed up to the concentration of 25 mg/L for 24 h. However, at 30 mg/L, 30% mortality was observed. Therefore, in present experiment the 27.5 mg/L concentration of imidacloprid for 24 h was used. For, dichlorvos, the reported concentration is 05 mg/L for acetyl cholinesterase inhibition study in zebrafish [13]. No mortality was found up to the concentration of 15 mg/L for 24 h. However, at 20 mg/L, 100% mortality was observed accordingly, 15 mg/L concentration was used in present experiment for dichlorvos. For atrazine, LC50 of atrazine in zebrafish is 9.56 mg/L for 96 h [14]. No mortality was observed up to the concentration of 03 mg/L for 24 h. However, at 04 mg/L, 50% mortality was observed in 24 h so, 03 mg/L concentration was used in present experiment for atrazine. For combined exposure, half concentrations of individual pesticides were used. The experiment was conducted in accordance with organisation for economic co-operation and development (OECD) guidelines.

Total 90 adults (4–5 months) zebrafish of length 2.8 ± 0.5 cm and weight 0.295–0.395 gms were procured from pet shop certified as wild species and handled in accordance with good animal practice as defined by the animal welfare bodies and the university committee No. PhD/FS/RA/02. Zebrafish were kept under controlled condition with temperature (25 ± 1 °C), pH 7.0 ± 0.2, conductance 0.2 μS cm−1, dissolved oxygen 7.2 ± 0.3 mg/L, hardness 110.0 mg/L, alkalinity 0.25 μS cm−1 and 12 h light and dark cycles. Animals were not fed 24 h prior to experimentation. The rationale for pesticides concentration was based on pilot experiments and literature. For, imidacloprid, no data is available for zebrafish. However, the dose of 10 mg/kg/day for rats was observed as no observed effect level (NOEL) [6]. Thus, an experiment was conducted and no mortality was observed up to the concentration of 25 mg/L for 24 h. However, at 30 mg/L, 30% mortality was observed. Therefore, in present experiment the 27.5 mg/L concentration of imidacloprid for 24 h was used. For, dichlorvos, the reported concentration is 05 mg/L for acetyl cholinesterase inhibition study in zebrafish [13]. No mortality was found up to the concentration of 15 mg/L for 24 h. However, at 20 mg/L, 100% mortality was observed accordingly, 15 mg/L concentration was used in present experiment for dichlorvos. For atrazine, LC50 of atrazine in zebrafish is 9.56 mg/L for 96 h [14]. No mortality was observed up to the concentration of 03 mg/L for 24 h. However, at 04 mg/L, 50% mortality was observed in 24 h so, 03 mg/L concentration was used in present experiment for atrazine. For combined exposure, half concentrations of individual pesticides were used. The experiment was conducted in accordance with organisation for economic co-operation and development (OECD) guidelines.

2.2. Animals, treatment and collection

Adult (4–5 months) zebrafish of length 2.8 ± 0.5 cm and weight 0.295–0.395 gms were procured from pet shop certified as wild species and handled in accordance with good animal practice as defined by the animal welfare bodies and the university committee No. PhD/FS/RA/02. Zebrafish were kept under controlled condition with temperature (25 ± 1 °C), pH 7.0 ± 0.2, conductance 0.2 μS cm−1, dissolved oxygen 7.2 ± 0.3 mg/L, hardness 110.0 mg/L, alkalinity 0.25 μS cm−1 and 12 h light and dark cycles. Animals were not fed 24 h prior to experimentation. The rationale for pesticides concentration was based on pilot experiments and literature. For, imidacloprid, no data is available for zebrafish. However, the dose of 10 mg/kg/day for rats was observed as no observed effect level (NOEL) [6]. Thus, an experiment was conducted and no mortality was observed up to the concentration of 25 mg/L for 24 h. However, at 30 mg/L, 30% mortality was observed. Therefore, in present experiment the 27.5 mg/L concentration of imidacloprid for 24 h was used. For, dichlorvos, the reported concentration is 05 mg/L for acetyl cholinesterase inhibition study in zebrafish [13]. No mortality was found up to the concentration of 15 mg/L for 24 h. However, at 20 mg/L, 100% mortality was observed accordingly, 15 mg/L concentration was used in present experiment for dichlorvos. For atrazine, LC50 of atrazine in zebrafish is 9.56 mg/L for 96 h [14]. No mortality was observed up to the concentration of 03 mg/L for 24 h. However, at 04 mg/L, 50% mortality was observed in 24 h so, 03 mg/L concentration was used in present experiment for atrazine. For combined exposure, half concentrations of individual pesticides were used. The experiment was conducted in accordance with organisation for economic co-operation and development (OECD) guidelines.

Total 90 adults (4–5 months) zebrafish were divided into five group (one control + four exposed) viz. Group: 1 (de-ionized water as control), Group: 2 (IMD, 27.5 mg/L), Group: 3 (DIC 15 mg/L), Group 4 (ATZ, 03 mg/L) and Group 5 (IMD 13.75 mg/L + DIC 7.5 mg/L + ATZ 1.5 mg/L) under semi-static conditions. At the end of the exposure period (24 h), the fish were sacrificed under a stereo microscope. Liver, kidney and brain were collected, minced and homogenized (2.5% w/v) with ice-cold 0.15% KCI-0.1 M phosphate buffer (pH 7.4).

2.3. Biochemical assays

An end product of lipid peroxidation is MDA (Malondialdehyde), were measured in tissue homogenates on the basis of the reaction with thiobarbituric acid (TBA) to form a pink colour complex, MDA produced was determined with the absorbance coefficient of the MDA-TBA complex at 550 nm using 1, 1, 3,3-tetraethoxypropane as the standard [18]. Glutathione levels were determined using 5,5’-dithio-bis (2-nitrobenzoic acid) (DTNB) for colour development at 420 nm. A standard curve using reduced glutathione was used for calibration [19]. The activity of superoxide dismutase was determined in the tissue homogenates by the modified method of NADH-phenazine methosulphate-nitroblue tetrazolium formazan inhibition reaction spectrophotometrically, measured at 550 nm [20]. The activity of catalase was determined at 550 nm [21]. The activity of glutathione peroxidase was determined using Glutathione as substrate and DTNB as standard at 420 nm [22].

2.4. Statistical analysis

Statistical significance of mean value of different biochemical parameters (LPO and GSH), antioxidant enzyme (SOD, CAT and GPx) in different tissue (liver, kidney and brain) for different groups (control, IMD, DIC, ATZ and CMD) tested by two-way analysis of variance (ANOVA).

3. Results

3.1. Lipid peroxidation

In liver, significant induction of MDA levels was observed in IMD, DIC and CMD group, while no change was observed in ATZ in compare to control (Fig. 1). In kidney, significant induction of MDA levels observed in the DIC and CMD group, while IMD and ATZ shows no change in compare to control. Significant alterations were not observed in MDA levels in brain of individually exposed groups. Interestingly, in combined exposure (at a half concentration of individual pesticides), the brain showed significant elevation in MDA levels in compare to control and all individually exposed group.

3.2. Glutathione

In liver, elevation of GSH content was observed in individually exposed IMD and DIC groups (Fig. 2). In contrast, GSH content depleted significantly in CMD group. In kidney, significant increase in GSH content observed in all three individually exposed groups. However, CMD group shows significant decrease in GSH content when compared with all individually exposed groups. Significant elevation observed in brain of IMD and DIC exposed fish however, no changes were observed in ATZ and CMD exposed group in compare to control group. Decrease in GSH content observed in CMD group when compared with IMD and DIC groups.

3.3. Superoxide dismutase activity

In liver, significant increase in SOD activity was observed in DIC and CMD group in compare to control, IMD and ATZ group (Fig. 3). No change was observed in ATZ group, while, IMD group shows decrease in SOD activity in compare to control. In kidney, significant decrease in SOD activity was observed in IMD and ATZ exposed groups in compare to control. In brain, depletion in SOD activity observed in IMD and DIC exposed groups in compare to control. While, no changes were observed in individual exposed ATZ group. Interestingly, in CMD exposed group (at a half dose of individual pesticide), the brain tissue showed significant elevation in SOD activity in compare to control and all individually exposed groups.
3.4. Catalase activity

In liver, significant depletion in CAT activity observed in IMD and ATZ groups, which significantly elevated in CMD group (Fig. 4). In kidney, no significant alterations were observed in IMD and ATZ exposed groups but, DIC group showed depletion. However, significant enhancement observed in CMD group in compare to control and individually exposed group. In brain, no significant alterations were observed in IMD and ATZ groups but DIC group shows depletion. However, significant enhancement was observed in CMD group in compare to control, and individually exposed group.

3.5. Glutathione peroxidase activity

Depletion in GPx activity observed in liver of IMD and ATZ exposed groups and no significant alteration was observed in DIC group in compare to control (Fig. 5). However, CMD group shows elevated activity in liver in compare to control and individually exposed groups.

Kidney shows depleted activity in IMD, but no changes were observed in DIC and ATZ group. However, CMD group shows elevation in compare to control and individually exposed groups. In brain, significant depleted activity observed in ATZ group, and no significant alterations were observed in IMD and DIC exposure. However, significant elevation was observed in CMD exposure in compare to control as well as individual exposure of pesticides.

4. Discussion

The biochemical alterations in different tissues of fish due to toxic effects of different pesticides have been reported extensively. Lipid peroxidation is the most important mechanism involved in pesticide induced toxicity [15]. Significant induction of MDA levels observed in liver and kidney of IMD and DIC group. Similar results were observed by Wang et al. [17] and El-Gendy et al. [23] due to imidacloprid exposure. Sukritha and Usharani [13] found elevated MDA levels in liver and kidney of dichlorvos exposed animal. Similarly, CMD group
shows highest alteration in comparison with control and exposed groups. Increased MDA levels suggest that an increase ROS in zebrafish. Due to pesticide exposure liver, kidney and brain are prone to oxidative stress and exhibit overall imbalance between oxidative stress and antioxidative capacity of cell.

Glutathione is an essential endogenous tripeptide, which prevents the cell from oxidative injury. It has antioxidative properties as contain thiol group, which act as a reducing agent however, GSH can be reversibly oxidized and reduced in cell. Our results of individual exposure are contrast with the findings of Kapoor et al. [6] and El-Gendy et al. [23] as they reported reduced GSH content. Interestingly, CMD group shows toxicity and produce depleted GSH levels which indicate the generation of ROS due to the exposure of the mixture of pesticides and to counteract ROS effects, more GSH consumed by related enzymes, this hypothesis also gets support from the elevated GPx activity in CMD group.

Antioxidant enzymes like SOD, CAT and GPx are the most important enzymes to maintain ROS balance as they scavenge the excess ROS and can be the first line of defence against ROS. SOD and CAT are the enzymes which convert superoxide (O2⁻) into H2O2 and then into H2O and O2. Our results are parallel with the findings of Kapoor et al. [6] who reported decrease in SOD activity due to IMD exposure in liver, kidney and brain. Decrease in activity of SOD may be due utilization of this enzyme in O2⁻ to H2O2. In DIC group, liver shows elevation in SOD activity which may be due to the fact that liver plays major role in the metabolism of xenobiotics. Our results are parallel with the findings of Celik and Suzek [1] who found elevated SOD activity due to DIC exposure. Findings of SOD are similar with the results of Blahova et al. [10] and Zhu et al. [16] who found elevated SOD activity in zebrafish due to Atrazine exposure. Interestingly, CMD group shows marked elevation in SOD activity in all the three examined tissue. However, in CMD group kidney shows elevation in SOD activity in comparison with individually exposed IMD group. Enhanced SOD activity reflects that the enzyme contributes to the elimination ROS induced by mixture of pesticides.

The primary work of catalase enzyme is to convert H2O2 into H2 and O2. Our results are analogous to the findings of Wang et al. [17] and El-Gendy et al. [23] as they documented elevated CAT activity due to imidacloprid exposure. In DIC and ATZ group decrease in CAT activity was observed our results are similar with the findings of Sukritha and Usharani [13]; Zhu et al. [16] and Blahova et al. [10]. However, CMD group shows marked elevation in CAT activity in comparison with control and exposed groups. The elevated enzyme activity might be due to the functioning of defence mechanism, which counteracts the oxidative stress induced by combination of said pesticides. Glutathione peroxidase is the general name of an enzyme family with peroxidase whose main biological role is to protect the organism from oxidative damage. GPx protects the cell against xenobiotic at the expense of GSH. The main function of GPx is to reduce lipid hydroperoxides to their consequent alcohols and to reduce free hydrogen peroxide El-Gendy et al. [23]. Significant decrease in GPx activity was observed in IMD and ATZ, our results are similar with the results of Kapoor et al. [6] for imidacloprid exposure. While, Blahova et al. [10] observed enhanced GPx activity due Atrazine exposure.

Plenty data is available for the presence of pesticides in river and other fresh water sources [12,24,25]. Since many types of pesticides coexist in ecosystem, and thus, present study will be helpful to understand the effects caused by mixture of pesticides.

5. Conclusion

From the present findings, it can be concluded that exposure to mixture of pesticides is more toxic in compare to individual pesticide, even at half concentration of individual pesticide. Individual pesticide shows specific mode of action, mechanism of toxicity and selected target organ, however in combination they may develop synergistic toxicity to any or all organs.

Conflict of interest

The authors declare that there is no conflict of interest. The research was conducted with no financial conflict or others factors which is considered to be declared as conflict.

Funding

This research did not receive any specific grant from funding agencies in the public, commercial, or not-for-profit sectors.

Acknowledgement

Saurabh Shukla is grateful to University Grants Commission (UGC), New Delhi for the award of Junior and Senior Research Fellowships.

References

[1] I. Celik, H. Suzek, Effects of sub acute exposure of dichlorvos at sublethal dosages on erythrocytes and tissue antioxidant defense systems and lipid peroxidation in rats, Ecotoxicol. Environ. Saf. 72 (2009) 955–958.
[2] R. Agarwal, S.K. Goel, J.R. Behari, Detoxification and antioxidant effects of curcumin in rats experimentally exposed to mercury, J. Appl. Toxicol. 30 (2010) 457–468, http://dx.doi.org/10.1002/jat.1517.
[3] K. Howe, M.D. Clark, C.F. Torroja, J. Torrance, C. Berthelot, M. Muffato, J.E. Collino, et al., The zebrafish reference genome sequence and its relationship to the human genome, Nature 496 (2013) 498–503.
[4] F. Melina, N. Garcia-Reyero, F. Padrius, P.J. Babin, S. David, C. Jerome, P. Eva, et al., Zebrafish models for human acute organophosphorus poisoning, Sci. Rep. 5 (2015) 15591, http://dx.doi.org/10.1038/srep15591.
[5] D.L. Chao, J.E. Caiba, Interaction of imidacloprid metabolites and analogs with the nicotinic acetylcholine receptor of mouse brain in relation to toxicity, Pest. Biochem. Physiol. 58 (1997) 77–88.
[6] U. Kapoor, M.K. Srivastava, S. Bharadwaj, L.P. Srivastava, Effect of imidacloprid on antioxidant enzymes and lipid peroxidation in female rats to derive its no observed effect level (NOEL), J. Toxicol. Sci. 35 (4) (2010) 577–581.
[7] P.J. Berry, F. Florence Buronfosse, B. Videmann, T. Buronfosset, Evaluation of the toxicity of imidacloprid in wild birds. A new high performance thin layer chromatography method for the analysis of liver and crop samples in suspected poisoning cases, J. Liq. Chromatogr. Relat. Techn. 22 (1999) 1547–1559, http://dx.doi.org/10.1081/JLC-100101750.
[8] ATSDR, Agency for Toxic Substances and Disease Registry Toxicological Profile for Dichlorvos, Public Health Service, US Department of Health and Human Services, Atlanta, GA, 1997.
[9] T.M. Bui-Nguyen, C.E. Baer, J.A. Lewis, D. Yang, P.J. Lein, D.A. Jackson, Dichlorvos exposure effects in large scale disruption of energy metabolism in the liver of the zebrafish Danio rerio, BMC Genom. 16 (1) (2015) 853, http://dx.doi.org/10.1186/s12864-015-1941-2.
[10] J. Blahova, L. Pihalova, M. Hostovsky, L. Divisova, R. Dobiskova, I. Mikulikova, S. Stepanova, Z. Svobodova, Oxidative stress responses in zebrafish (Danio rerio) after subchronic exposure to Atrazine, Food Chem. Toxicol. 61 (2013) 82–85, http://dx.doi.org/10.1016/j.fct.2013.02.041.
[11] M.E. Andersen, J.E. Dennison, Mechanistic approaches for mixture risk assessments: present capabilities with simple mixtures and future directions, Environ. Toxicol. Pharmacol. 16 (1–2) (2004) 1–11.
[12] A. Cranecapa, A. Manis, A. Navarro-Ortega, Y. Picó, D. Barceló, Pesticides in the Ebro River Basin: occurrence and risk assessment, Environ. Pollut. 231 (2016) 14–24, http://dx.doi.org/10.1016/j.envpol.2015.12.059.
[13] T.H. Sukritha, M.V. Usharani, Effects of organophosphates on acute poisoning and acetylcholinesterase inhibition in zebrafish, Int. J. Bioanalyses 2 (3) (2013) 575–580.
[14] G.A. Abeer, Y.Y. Al-Sawafi, Bioconcentration and antioxidant status responses in zebrafish (Danio rerio) under atrazine exposure, Int. J. Chem. Eng. Appl. 4 (2013) 204–208, http://dx.doi.org/10.1155/2013/V4.295.
[15] C.H.D. Nwan, W.I. Laka, N.S. Nagurue, R. Kumar, B. Kushwaha, S.K. Srivastava, Toxicity of the herbicide atrazine: effects on lipid peroxidation and activities of antioxidant enzymes in the freshwater fish channa punctatus (Bloch), Int. J. Environ. Res. Public Health 7 (2010) 3298–3312.
[16] L.S. Zhu, B. Shao, Y. Song, H. Xie, J.H. Wang, W. Liu, X.X. Hou, DNA damage and effects on antioxidant enzymes in zebra fish (Danio rerio) induced by Atrazine, Toxicol. Mech. Meth. 21 (1) (2011) 31–36, http://dx.doi.org/10.3109/15376516. 2011.529186.
[17] J. Wang, W. Ge, S. Yan, L. Zhu, Oxidative stress and DNA damage induced by imidacloprid in zebrafish (Danio rerio), J. Agric. Food Chem. 63 (6) (2015) 1856–1862.
[18] H. Okhawa, N. Oishi, K. Yaicy, Assay for lipid peroxidation in animal tissues by thiobarbituric acid reaction, Annu. Biochem. 95 (1979) 351–358.
[19] G.L. Ellman, Tissue sulfhydryl groups, Arch. Biochem. Biophys. 82 (1959) 70–77.
[20] P. Kakkad, B. Das, P.N. Vishwanathan, A modified spectrophotometric assay of
superoxide dismutase, Ind. J. Biochem. Biophys. 21 (1984) 130–132.
[21] A.K. Sinha, Colorimetric assay of catalase, Anal. Biochem. 47 (1972) 389–394.
[22] L. Flohe, W.A. Gunzler, Assays of glutathione peroxidize, Meth. Enzymol. 105 (1984) 114–121.
[23] K.S. El-Gendy, N.M. Aly, F.H. Mahmoud, A. Kenawy, A.K. El-Sebae, The role of vitamin C as antioxidant in protection of oxidative stress induced by imidacloprid, Food Chem. Toxicol. 48 (2010) 215–221.

[24] V.C. Schreiner, E. Szöcs, A.K. Bhowmik, M.G. Vijver, R.B. Schäfer, Pesticide mixtures in streams of several European countries and the USA, Sci. Total Environ. 573 (2016) 680–689, http://dx.doi.org/10.1016/j.scitotenv.2016.08.163.
[25] Y. Wang, L. Lv, Y. Yu, G. Yang, Z. Xu, Q. Wang, L. Cai, Single and joint toxic effects of five selected pesticides on the early life stages of zebrafish (Denio rerio), Chemosphere 170 (2017) 61–67, http://dx.doi.org/10.1016/j.chemosphere.2016.12.025.