Rapid formation of isoprene photo-oxidation products observed in Amazonia

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Abstract. Isoprene represents the single most important reactive hydrocarbon for atmospheric chemistry in the tropical atmosphere. It plays a central role in global and regional atmospheric chemistry and possible climate feedbacks. Photo-oxidation of primary hydrocarbons (e.g. isoprene) leads to the formation of oxygenated VOCs (OVOCs). The evolution of these intermediates affects the oxidative capacity of the atmosphere (by reacting with OH) and can contribute to secondary aerosol formation, a poorly understood process. An accurate and quantitative understanding of VOC oxidation processes is needed for model simulations of regional air quality and global climate. Based on field measurements conducted during the Amazonian Aerosol Characterization Experiment (AMAZE-08) we show that the production of certain OVOCs (e.g. hydroxyacetone) from isoprene photo-oxidation in the lower atmosphere is significantly underpredicted by standard chemistry schemes. Recently reported fast secondary production could explain 50% of the observed discrepancy with the remaining part possibly produced via a novel primary production channel, which has been proposed theoretically. The observations of OVOCs are also used to test a recently proposed HOx recycling mechanism via degradation of isoprene peroxy radicals. If generalized our observations suggest that prompt photochemical formation of OVOCs and other uncertainties in VOC oxidation schemes could result in uncertainties of modelled OH reactivity, potentially explaining a fraction of the missing OH sink over forests which has previously been largely attributed to a missing source of primary biogenic VOCs.

1 Introduction

Volatile organic compounds (VOCs) critically influence the composition of the Earth’s atmosphere by fueling tropospheric chemistry (Atkinson, 2000), and providing condensible oxidation products for organic aerosol formation (Kanakidou et al., 2005). On a global scale the emission strength of biogenic VOCs dominates the annual VOC budget (~1000–2000 Tg/y). A large fraction of this reduced carbon flux enters the atmosphere in form of isoprene, which could exceed global methane emissions (Guenther et al., 2006). VOCs are emitted from many terrestrial plants at high rates, in particular in tropical ecoregions, and some species are known to re-emit up to 10% of their assimilated carbon in form of isoprene (Kesselmeier et al., 2002). Isoprene is also highly reactive and its photochemical evolution therefore plays a central role in atmospheric chemistry.

Detailed chemical schemes are needed in global atmospheric chemistry models (e.g. Brasseur et al., 1998; Bey et al., 2001; Kuhlmann et al., 2003) to simulate the tropical photo-reactor and assess how the oxidizing capacity of the remote tropical atmosphere is modulated by this compound. Isoprene chemistry schemes (Fan and Zhang, 2004; Carter and Atkinson, 1996; Pinho et al., 2005) are typically condensed so they can be incorporated in global and regional chemistry models. Most current lumped chemistry schemes include near explicit representation of the first and second generation oxidation products of isoprene (Emmons et al., 2009; Taraborrelli et al., 2008; Emmerson and Evans, 2009). Some of these schemes have been applied in field studies conducted in the tropics (e.g. Warneke et al., 2001; Ganzeveld et al., 2008). Recent theoretical and laboratory evidence suggested that some of the basic steps of isoprene
peroxy radical cycling are still poorly understood (Dibble, 2004; Park et al., 2003; Peeters et al., 2009; Paulot et al., 2009a) and ramifications for \(a\)keri cycling have been discussed (Lelieveld et al., 2008; Hofzumahaus et al., 2009; Paulot et al., 2009a). Here we use field measurements of isoprene and its oxidation products methyl vinyl ketone (MVK), methacrolein (MAC) and hydroxycetone to assess their photochemical evolution and test existing kinetic schemes at a remote field site approx. 60 km NNW of Manaus in the central Amazon basin during the wet season in 2008.

2 Methods

2.1 Measurement site

Measurements were conducted as part of the Amazonian Aerosol Characterization Experiment (AMAZE-08) from 9 to 28 February 2008. The site (02°35.65°S, 60°12.55°W) located in the Reserva Biologica do Cuieiras and managed by the Instituto Nacional de Pesquisas da Amazonia (INPA) and the Large-Scale Biosphere-Atmosphere Experiment in Amazonia (LBA). The vegetation cover consists of primary tropical rainforest (approx. leaf area index of 5–6 m²/m²) with an average canopy height of ~30 m surrounding the ~40 m measurement tower.

2.2 VOC measurement and data analysis

A Proton-Transfer-Reaction Mass Spectrometer was used for gradient measurements of selected VOCs. The instrument sequentially sampled six independent 1” inch Teflon (PFA) sampling lines mounted at 2, 10.9, 16.7, 23.9, 30.3, and 39.8 m on the tower. A valve switching system changed sampling lines every 5 min and cycled through the entire gradient over a 30 min period. Gradients were calculated from the 5 min averages. High flow rates through the sampling lines resulted in delay times of less than 8–12 s, measured by spiking a VOC pulse at each sampling inlet.

Isoprene and the sum of MVK and MAC were measured at ion channels \(m/z \) 69 and \(m/z \) 71 respectively. The specificity of the PTRMS to monitor isoprene and the sum of MVK+MAC has been demonstrated previously (e.g. de Gouw and Warneke, 2007). Hydroxycetone was monitored on ion channel \(m/z \) 75 and was identified as a major VOC observed by PTRMS in tropical ecosystems in the past (Williams et al., 2001). The detection limits for isoprene, MVK+MAC and hydroxycetone for a 5 s integration time were 10.5 and 5 pptv respectively. Particular sensitivities corresponded to proton-transfer reaction rate constants of 2e-9, 3e-9 and 3.5e-9 cm³/s respectively. PTRMS hydroxycetone concentration data from earlier field campaigns (Williams et al., 2001; Holzinger et al., 2007; Stroud et al., 2005) were corrected by the relative difference of the rate constants between hydroxycetone and isoprene. Potential interferences on \(m/z \) 75 due to butanol can be excluded as butanol completely dehydrates and is observed exclusively on \(m/z \) 75 under operating conditions used here. Propionic acid also dehydrates partially into \(m/z \) 57. Our dataset shows no significant correlation between signals observed on \(m/z \) 57 and \(m/z \) 75, which is indicative that propionic acid did not contribute to \(m/z \) 75. The most likely candidate for \(m/z \) 75 is therefore a compound originating from isoprene oxidation. We can further exclude other interferences such as diethyl ether and methyl acetate from GC samples. We conclude that the PTRMS mass channel \(m/z \) 75 was specific for the measurement of hydroxycetone in this high isoprene environment and confirm conclusions drawn from earlier studies (Williams et al., 2001).

GC-MS air samples were collected on carbotrap and tenax cartridges that were stored at approximately 0°C until analysis at NCAR Boulder laboratory, except during transit from Brazil to USA, when they were at ambient temperature for approximately 1 day. The cartridges were desorbed thermally using an NCAR-made system (Greenberg et al., 1994) and analyzed by gas chromatography with mass spectrometric detection (Hapsite Smart, Inficon, East Syracuse NY) using a 30 m × 0.3 mm ID 1 mm film DB-1 column, temperature programmed from 40 to 200°C at 3°C per min after an initial 2 min hold. VOC were quantified with respect to NIST traceable standards as described by Greenberg et al. (1994).

Ozone (\(O_3\)), and nitric oxide (NO) concentrations were also measured via the 6-level sampling manifold. Ozone was measured by UV absorbance (2B Technologies, Model 205) every 10 s and then averaged over the entire 5 min sampling time on each level, excluding only the first 15 s to insure adequate flushing of the connecting gas lines. The ozone analyzer was compared with laboratory instruments both prior to and following the experiment and found to agree with ±5% with a detection limit of 2 ppbv. It was zeroed periodically by placing an ozone scrubber on the 1.5 m inlet. NO was measured by ozone-induced chemiluminescence (Ecophysics, Model 88Y) at a sample rate of 1 s. Measurements were averaged over the first 2 min of sampling
on a given level (excluding initial 15 s). During the other 3 min, the air was either passed through a photolytic converter (Droplet Technology, BLC-100) to convert NO₂ to NO or through a molybdenum catalyst which reduces oxidized nitrogen to NO which was subsequently detected by chemiluminescence. The NO analyzer was zeroed automatically every hour by diverting the sample flow through a pre-reactor and adding a large excess of ozone which consumed the NO. It was calibrated periodically during the campaign by standard addition of a known NO concentration to the 2m inlet. The detection limit was ∼50 pptv. Laboratory tests indicated no measurable loss of either ozone, NO or NO₂ within any of the PFA inlet lines. Wind velocities at 40 m were measured by a 3-dimensional sonic anemometer (Applied Technologies, SATI-K) mounted on a boom extending 2 m from the tower and sampled at 20 Hz.

2.3 VOC gradient fluxes

Ecosystem-scale fluxes of VOCs were calculated based on concentration gradients throughout the canopy and applying an Inverse Lagrangian Transport Model (Raupach, 1989, Nemitz et al., 2000, and Karl et al., 2004). The VOC fluxes were computed according to \( \vec{C} - \vec{C}_{ref} = \vec{D} \cdot \vec{S} \), where \( \vec{C} \) is the VOC concentration (\( \mu g/m^3 \)) vector for each level, \( \vec{C}_{ref} \) is the VOC concentration (\( \mu g/m^3 \)) at reference height (e.g. 14 m), \( \vec{D} \) (m) represents a dispersion matrix and \( \vec{S} \) (mg/m³/h/m) the resulting VOC source/sink vector. \( \vec{D} \) can be expressed as a function of Lagrangian timescale (TI) and profiles of the standard deviation of the vertical wind speed (\( \sigma_w \)) divided by the friction velocity (\( u_* \)). The calculation was performed using a 10×4 dispersion matrix. Integration over all source and sink terms (S) yielded the canopy scale VOC flux (mg/m²/h). Fluxes were calculated for 30 min intervals. The parameterization of \( \vec{D} \) was based on turbulence measurements inside and above the canopy and calculated using the far- and near-field approach described by Raupach (1989). The Lagrangian timescale was parameterized according to Raupach (1989).

3 Results and discussion

3.1 Biosphere-Atmosphere exchange of isoprene, MVK+MAC and Hydroxyacetone

Figure 1 (left panel) shows the exchange of VOCs at the atmosphere-biosphere interface and depicts the mean integrated vertical source/sink distribution of isoprene, MVK, MAC and hydroxyacetone measured by PTR-MS during daytime (11:00–17:00 local time (LT)). The source/sink distribution was inferred from in-canopy concentration measurements using an inverse Lagrangian transport model. Average midday ozone mixing ratios (Fig. 1, right panel) were ∼8 ppbv decreasing to <3 ppbv near the ground. Isoprene was mostly emitted in the upper part of the canopy as expected because of the light environment and the distribution of sun and shade leaves throughout the canopy. The oxidation products MVK, MAC and hydroxyacetone were
of these VOCs suggests that the primary source was located above the forest. The analysis shows that primary biological emissions of MVK, MAC and hydroxyacetone were negligible during this study.

Observed OVOC mixing ratios are determined by the balance of their combined loss and production rates. While their production rates are dominated by photochemistry, their loss rates are influenced by photochemical destruction, dry deposition and vertical mixing (such as entrainment of air from the free troposphere). In order to estimate the relative importance of these three loss terms we assume a 1000 m deep well-mixed planetary boundary layer (PBL) (Karl et al., 2007). The lifetime due to dry deposition equals the mixed layer concentration times the PBL height divided by the deposition flux (Fig. 1). Taking mixing ratios of 3 and 1.5 ppbv for MVK+MAC and hydroxyacetone, the lifetime due to dry deposition is on the order of 25 and 22 h respectively. Similarly the influence of entrainment can be estimated. Mixing ratios in the free troposphere (FT) are typically an order of magnitude lower (Karl et al., 2007; Kuhn et al., 2007). The concentration jump of a VOC between the PBL and FT can then be approximated by the mixed layer concentration. For the measured entrainment velocities obtained during an earlier study (e.g. Karl et al., 2007; 8–10 cm/s) the dilution timescale of vertical mixing is 3–4 h.

These physical loss processes can be compared to the chemical destruction rate for a daytime OH density of 5e6 molecules/cm³ (Ganeveld et al., 2008). The corresponding chemical lifetimes for isoprene, MVK+MAC and hydroxyacetone are 0.5, 2 and 18 h respectively. Since daytime concentrations of isoprene, MVK, MAC and hydroxyacetone in the FT are typically much lower than in the PBL the effect of vertical dilution has minimal influence on VOC correlations observed in the PBL, because all VOCs are diluted at the similar rates. When investigating correlations of OVOCs (e.g. MVK and MAC) with respect to isoprene, the additional sink due to OVOC dry deposition would result in systemically lower OVOC mixing ratios than would be calculated from photochemical destruction alone.

### 3.2 Photochemical production of MVK, MAC and hydroxyacetone

The first steps of the photo oxidation of isoprene via the hydroxyl radical (OH) can be summarized as:

- \( {\text{C}_5\text{H}_8 + \text{OH}} \overset{O_2}{\rightarrow} {\text{C}_5\text{H}_9O_3} \) (R1a)
- \( {\text{C}_5\text{H}_9O_3 + \text{NO} \rightarrow \text{carbonyls} } \) (R1b)
- \( {\text{C}_5\text{H}_9O_3 + \text{HO}_2 \rightarrow \text{peroxides} } \) (R1c)

Measured NO mixing ratios were sufficiently high (100–300 pptv, lifetime \( \tau \approx 30 \) s) in the surface layer to dominate over the reaction with \( \text{HO}_2 \) radical estimates (10–20 pptv; \( \tau \approx 30 \) s) according to results from a photochemical box model based on the Mozart Chemistry Mechanism v4 (Emmons et al., 2009). Consequently peroxy radical reactions with NO (\( \tau \approx 30 \) s) led to significant production of the carbonyls methyl vinyl ketone (MVK) and methacrolein (MAC). For VOCs discussed here Reaction (R1) can therefore be simplified to the following kinetic reaction model, where \( Y \) is the branching ratio for each channel.

- \( {\text{C}_5\text{H}_8 + \text{OH}} \overset{0.32}{Y} \rightarrow {\text{MVK} } \) (R2a)
- \( {\text{C}_5\text{H}_8 + \text{OH}} \overset{0.32}{Y} \rightarrow {\text{MAC} } \) (R2b)
- \( {\text{MAC} + \text{OH}} \overset{0.42}{Y} \rightarrow \text{hydroxyacetone} \) (R2c)
- \( {\text{MVK} + \text{OH}} \rightarrow \text{Products} \) (R2d)

Uncertainties for yields are typically on the order of 20% (Atkinson et al., 1997). While the focus in this section is on the observed mixing ratios of isoprene, MVK, MAC and hydroxyacetone a more detailed representation of the atmospheric production of OVOCs from isoprene is discussed in Sect. 3.3. During daytime the evolution of isoprene chemistry in the PBL is conceptually different than many environmental chamber experiments in that there is continuous supply of isoprene. As a result the production of compounds originating directly from first generation isoprene peroxy radicals is relatively more prominent in the real atmosphere. For example the ratio of (MVK+MAC)/isoprene will approach infinity in batch-mode environmental chambers, while there is a steady-state upper limit of \( \sim 2.4 \) in the PBL (assuming yields according to reaction R2 and OH reaction rate constants of \( 1e-10 \) cm/s, \( 1.9e-11 \) cm/s and \( 3.4e-11 \) cm/s for isoprene, MVK and MAC respectively). However, due to the time-dependent isoprene supply and the slower OH reactions of the products, steady state conditions are rarely reached. During AMAZE-08 a daytime (11:00–17:00LT) ratio of 0.44±0.05 was observed (Fig. 2, left panel). Similar ratios were observed in previous field studies (e.g. Warneke et al., 2001; Kuhn et al., 2007; Eerdekens et al., 2008; Karl et al., 2007).

The major source of hydroxyacetone is oxidation of MAC (e.g. Carter and Atkinson, 1996). Figure 2 (right panel) shows a correlation plot between hydroxyacetone and the sum of the first generation oxidation products MVK and MAC. For a given MVK+MAC to isoprene ratio of 0.44 we calculate a slope of 0.03 between hydroxyacetone and MVK+MAC according to Reaction (R2) (assuming constant isoprene emissions). This is \( \sim 10 \) times smaller than the observed ratio (0.30±0.08) and implies that an additional source of hydroxyacetone, which is highly correlated with the oxidation products MVK and MAC (\( R^2=0.79 \)) and isoprene (\( R^2=0.70 \)), is missing. The high degree of correlation suggests that the missing source of hydroxyacetone comes...
Fig. 2. Left panel: Correlation between MVK+MAC and isoprene. Right panel: Correlation between hydroxyacetone and MVK+MAC. Regressions based on a robust fitting procedure are indicated by solid red lines (Holland et al., 1977). The dashed red lines indicate steady state limits. The modelled hydroxyacetone/(MVK+MAC) slope is depicted by the blue dashed line.

from within the isoprene oxidation chain, but has to be faster than the production from MAC, which is thought to be the major contributor to the secondary formation of hydroxyacetone. Other possible chemical sources of hydroxyacetone include oxidation of peroxides originating from propene and production from acetone oxidation (e.g. acetonyl peroxy radicals reacting with methyl peroxy radicals). However, both of these are expected to be slow due to the low concentrations of propene and subsequent peroxides or the need for radical-radical reactions in the case of acetone oxidation. Since isoprene mixing ratios (e.g. up to 8 ppbv) dominate the tropical environment these processes must compete with the dominant production path via MAC. Although anticipated to be small in the wet season of the central Amazon Basin, another possible source of hydroxyacetone is biomass burning (Yokelson et al., 2007). We investigated this using acetonitrile as specific gas phase marker. Periods of elevated acetonitrile (∼10% of the entire dataset) were excluded from the analysis. A further argument against a significant biomass burning source is the good correlation observed between hydroxyacetone and MVK+MAC as well as with isoprene.

In order to check whether the observed ratios between isoprene, MVK+MAC and hydroxyacetone are comparable to previous datasets we have re-analyzed data reported in the literature. The triangular plot shown in Fig. 3 summarizes findings from five field campaigns, including the present study, conducted near isoprene emission sources (Williams et al., 2001; Holzinger et al., 2007; Stroud et al., 2005; Spaulding et al., 2003). Concentrations from these datasets, which are representative for typical local noontime conditions, were normalized by the sum of isoprene, MVK, MAC and hydroxyacetone in order to quantitatively compare their relative ratios. The color coded triangular symbols represent a chemical trajectory along which the relative ratios would evolve according to Reaction (R2). The color coding indicates photochemical age (time exposure to OH). Due to a constant supply of isoprene the theoretical trajectory ends at the steady state limit, typically reached after 10⁷ molecules cm⁻³ days of OH exposure for the reaction sequence discussed here. Measured distributions from all field observations, in particular the AMAZE-08 dataset discussed here, show that hydroxyacetone mixing ratios are significantly higher than what would be expected from the reaction model described by Reaction (R2) when only production from MAC is assumed. The discrepancy becomes larger for data collected closer to the source corresponding to a recent supply of isoprene.

Reconciliation with such fast production of hydroxyacetone requires additional production paths from isoprene (e.g. through isoprene alkenoxy radicals). We extended the set of reactions listed in Reaction (R2) by including direct production from isoprene plus OH and obtained an optimal yield (Yiso) based on a non-linear least square regression procedure (Seber et al., 1989) of the coupled set of differential reactions (R2). The color coded circles in Fig. 3 depicts
Fig. 3. Triangular correlation plot for the measured and modelled mixing ratios of isoprene, MVK+MAC and hydroxyacetone. Mixing ratios are normalized by the sum of the concentrations of isoprene, MVK, MAC and hydroxyacetone. Modelled concentrations are color coded by photochemical age. Field datasets are depicted by symbols as following: red circle (AMAZE, 2008; this study), blue diamond (LBA-Claire, 1999; Williams et al., 2001), green triangle (CELTIC 2003; Stroud et al., 2005), magenta triangle (ICARTT, 2004; Holzinger et al., 2007) and black square (Blodgett Forest, 2003; Spaulding et al., 2003). Triangles represent the standard model (i.e. no fast production). Circles show the model case that includes fast production based on non-linear regression. The chemical trajectories of the models terminate at the steady state limit, typically reached after \(10^7\) molecules cm\(^{-3}\) days of OH exposure.

a curve based on direct production with a fitted Yiso of 8.3 (±2.2) %, which would be sufficient to explain the observed distribution of hydroxyacetone in isoprene dominated environments. Several studies have proposed primary (Dibble et al., 2004) and fast secondary production of hydroxyacetone (Paulot et al., 2009b). Dibble et al. (2004) recently conducted a theoretical study on the decomposition of certain isoprene – alkenoxy radicals and hypothesized a direct channel which would lead to the formation of products typically associated with secondary and tertiary chemistry (e.g. hydroxyacetone and glyoxal). Paulot et al. (2009b) reported very fast production of hydroxyacetone from isoprene nitrates and hydroxycarbonyls after initial reaction with OH.

In order to investigate the sensitivity of the proposed additional production paths for hydroxyacetone, we compared four reaction mechanisms. We implemented (1) the Mozart mechanism (Emmons et al., 2009), (2) a recently developed isoprene oxidation scheme proposed by Paulot et al. (2009b), which was extended for low NO\(_x\) conditions (Paulot et al., 2009a), (3) a simple sequential reaction model (SRM) summarized in Table 2 (Appendix), which was extended to include fast production of hydroxyacetone, and (4) a novel reaction mechanism for isoprene proposed by Peeters et al. (2009) (Table 3, Appendix). Figure 4 shows the results for these model runs, for the conditions encountered during the AMAZE-08 study (viz. constant isoprene emissions and NO mixing ratios of 100 pptv). The top panel in Fig. 4 shows the evolution of the ratio of MVK+MAC to isoprene. Observations during AMAZE-08 (Fig. 2) correspond to a ratio of 0.44, which is shown on the left side of Fig. 4.

The SRM, Mozart and Paulot mechanisms reproduce the observed MVK/MAC ratios well. The mechanism by Peeters et al. (2009), would predict a ratio of \(\sim 10\) for an observed (MVK+MAC) to isoprene ratio of 0.44 (see red vertical line in Fig. 4). Possible reasons for such a large discrepancy are explored in the next section.

The modelled and observed hydroxyacetone to (MVK+MAC) ratios are plotted in the lower panel of
Fig. 4. Observed (left edge) and modelled (main figure) ratios of OVOC to isoprene for four different models (Peeters et al., 2009; Paulot et al., 2009a and b; SRM, a sequential reaction model according to Table A2; Mozart, Emmons et al., 2009). Top: MVK+MAC/isoprene. Middle: MVK/MAC. Bottom: Hydroxyacetone/(MVK+MAC). The dashed black horizontal line corresponds to the (MVK+MAC)/isoprene ratio of 0.44 measured during AMAZE-08, and the dashed black vertical line shows the value of the corresponding photochemical age obtained by the predictions from three models (Paulot, SRM, Mozart). The vertical red line corresponds to the photochemical age by the mechanism of Peeters et al. (2009).

Fig. 4. For a given (MVK+MAC) to isoprene ratio of 0.44 (e.g. vertical black line), the modelled hydroxyacetone to (MVK+MAC) ratio obtained from the Paulot mechanism is closest to the observed ratio (e.g. within 50%), which was used to derive the modelled SRM curve. This suggests that fast secondary production of hydroxyacetone could explain about 50% of the observed hydroxyacetone to (MVK+MAC) ratio during AMAZE-08. The remaining 50% could be related to primary production mechanisms similar to those proposed by Dibble et al. (2004).

3.3 Implications for atmospheric chemistry

3.3.1 HOx recycling

In the last section OVOC ratios were compared to different photochemical oxidation mechanism. While three of these oxidation schemes could reproduce the observed MVK and MAC mixing ratios, the mechanism presented by Peeters et al. (2009) suffered from unrealistically high MVK/MAC ratios and underestimated the sum of MVK and MAC relative to isoprene. This mechanism is proposed as a novel HOx recycling scheme which could potentially explain large missing HOx sources in isoprene dominated environments reported by Lelieveld et al. (2008) and Hofzumahaus et al. (2009). The mechanism is summarized in the Table 3 in the Appendix.

Briefly, after initial addition of OH to one of the four unsaturated C-atoms of isoprene, O2 addition leads to the formation of three isoprene peroxyradicals, which are considered in thermodynamic equilibrium. The two major adducts (1-OH-isoprene and 4-OH-isoprene) account for 90% of the isoprene oxidation channels and are considered here. The Z-1-OH-4-OO* and Z-4-OH-1-OO* peroxyradicals are proposed to undergo a fast 1,6-H-shift (HO2 regenerating channels) that competes with conventional reaction channels through HO2, RO2 and NO. Similarly, the 1-OH-2-OO* and 4-OH-3-OO* peroxyradicals would undergo a 1,5-H-shift (OH regenerating channels) competing with reaction via HO2, RO2 and NO. These proposed radicals would then decompose in a single concerted reaction to generate HO2, OH, a hydroperoxymethylbutenal, MVK and MAC. The decomposition rates for the 1,6-H-shift (HO2 regenerating channels) that competes with conventional reaction channels through HO2, RO2 and NO. Similarly, the 1-OH-2-OO* and 4-OH-3-OO* peroxyradicals would undergo a 1,5-H-shift (HO2 regenerating channels) competing with reaction via HO2, RO2 and NO. These proposed radicals would then decompose in a single concerted reaction to generate HO2, OH, a hydroperoxymethylbutenal, MVK and MAC. The decomposition rates for the 1,6-H-shift reactions are highly dependent on energy barriers and are proposed to be on the order of 1 s−1 and 8 s−1. For comparison the reaction rates via the conventional NO channel would correspond to 0.01–0.03 s−1 for conditions of AMAZE-08. Thus one reason for the different MVK, MAC and hydroxyacetone ratios discussed earlier can be found in these fast decomposition rates.
Figure 5 shows the sensitivity of MVK, MAC and hydroxyacetone with respect to the 1,5-H-shift, 1,6-H-shift and reverse reaction rates of the Z-1-OH-4-OO* and Z-4-OH-1-OO* peroxyradicals. For example, a change in the reverse reaction rate (kr (Z-1-OH-4) - green dashed line in upper panel) of 100% leads to a change in the MVK concentration of 62% in steady state (e.g. >1e5 molecules cm^{-3} days). Similarly, a change of the rate constant of the 1,6-H-shift of the Z-1-OH-4 reaction (blue line) by 100% shifts the MVK concentration by ~59%. By changing the rate of the 1,6-H-shift for the Z-4-OH-1-peroxy radical by a factor of 10 (e.g. 0.8 s^{-1} vs 8 s^{-1}), the MAC concentration would increase by a factor of 5 and the hydroxyacetone concentration would increase between a factor of 8 at low photochemical age (<5e4 molecules cm^{-3} days) and 5 at high photochemical age (>1e6 molecules cm^{-3} days), where the production of hydroxyacetone is dominated by secondary production via MAC.

Reconciliation of the mechanism of Peeters et al. (2009) with our observed OVOC ratios would require that the 1,5-H and 1,6-H-shift decomposition reactions be as fast as the conventional reaction channels via NO and the reverse reaction rates of the Z-1-OH-4-OO* and Z-4-OH-1-OO* peroxyradicals be reduced. The implication is that the thermodynamic stability of these radicals would have to increase. Independent experimental evidence was recently presented by Paulot et al. (2009a), who tentatively concluded that much smaller yields for the (2Z)-hydroperoxymethylbutenal (e.g. <10%) than would be predicted via the 1,6-H shifts were needed to explain their observations. One consequence of these independent experimental results is that they suggest a lower HOx recycling efficiency. For the standard mechanism proposed by Peeters et al. (2009) we calculate a HOx yield of 0.67 (0.64 HO2+0.03 OH) for AMAZE-08 conditions. When adjusting reaction rates to match the AMAZE-08 OVOC observations we calculate an overall HOx yield of 0.22 (0.10 HO2+0.12 OH), corresponding to approximately 33% of the original yield.

3.3.2 Interpretation of OH reactivity measurements in isoprene dominated environments

If generalized these findings also have implications for OH reactivity measurements (di Carlo et al., 2004; Sinha et al., 2008) conducted close to isoprene emission sources. OH is considered the main oxidant in the atmosphere and reacts with all VOCs. The reaction frequency of a VOC with OH is defined as the product of its rate-coefficient times its concentration. The sum of all reaction frequencies is the inverse of the OH lifetime and is termed OH reactivity. The measurement of OH reactivity has been used as a top-down constraint to infer VOC sources in the atmosphere (e.g. di Carlo et al., 2004; Sinha et al., 2008). However, since all reactive VOCs are not always measured concurrently,
photochemical models must be used to estimate contributions of many species to the overall OH reactivity. Rapid formation of isoprene oxidation products could lead to a higher modelled OH reactivity than assumed based on current photo-oxidation mechanisms. This becomes particularly important when lumped chemical schemes are used (di Carlo et al., 2004), which further simplify the chemical degradation mechanism of VOCs.

In order to test the uncertainty of modelled OH reactivity, we compare four different oxidation schemes for the most prominent OVOCs produced from isoprene (Fig. 6). Figure 6a shows the evolution of OVOCs relative to isoprene inferred from the mechanism proposed by Paulot et al. (2009a, b). It includes the production of a newly discovered class of OVOCs (epoxides: IEPOX). For simplicity we have excluded the contribution from isoprene nitrates which contribute less than 2% for conditions encountered during AMAZE-08. Panel b depicts the OVOC evolution based on an extensive set of reactions implemented in a kinetic reaction model (see Tables 1 and 2 Appendix). The black dashed line shows the evolution of the OH reactivity of isoprene oxidation products without the fast OVOC production channels. Panel c shows results obtained from the Mozart mechanism (Emmons et al., 2009). As in panel b the black dashed line shows the model run without fast production of OVOCs. Panel d depicts the formation of isoprene photo-oxidation products based on the Regional Atmospheric Chemistry Mechanism (RACM) (Stockwell et al., 1997), which has previously been used to model OH reactivity for studies of missing OH sinks (di Carlo et al., 2004). Most of the reactivity in RACM is lumped into six species: formaldehyde, methacrolein, OLT (terminal alkenes, such as propane), methylglyoxal, hydroxyacetone and a generic ketone.

As discussed in Sect. 3.1 the influence of mixing is neglected because OVOC concentrations are assumed to be small in the FT compared to the PBL during daytime. This simplification results in a systematic overestimation of the OVOC reactivity, as they are longer-lived than isoprene and the concentration jump between the FT and the PBL would be smaller than for isoprene. Previous aircraft measurements have shown that MVK and MAC decrease rapidly in the FT similar to isoprene (Karl et al., 2007; Kuhn et al., 2007). Therefore the influence of vertical entrainment would likely impact the distribution on timescales longer than 5–6 h. For example for midday OH concentrations of 5e6 molecules/cm^3 vertical entrainment would mostly influence the reactivity distribution for photochemical ages greater than 1e6 molecules/cm^3 days.

For typical conditions during AMAZE-08 the modelled OH reactivity from photo-oxidation products could account for 20–150% of the measured reactivity relative to isoprene depending on the photochemical age. Formaldehyde (HCHO), an important intermediate isoprene oxidation product, can account for up to 30% of the OVOC reactivity. Together the major oxidation products MVK, MAC and HCHO
typically account for up to 75% of the OVOC reactivity. Adding rapid formation of OVOCs to this scheme could increase the reactivity of photo-oxidation products relative to isoprene by up to 20% for the kinetic reaction model (e.g., dashed line panel b). For the longest photochemical age considered (5e6 molecules/cm$^3$ days) the maximum contribution of the OVOC reactivity relative to isoprene between the Paulot mechanism (panel a), the sequential reaction model (panel b) and the Mozart mechanism (panel c) falls between 2 and 2.7. The RACM mechanism would predict a ratio of 1.5. For shorter photochemical ages the range of modelled OVOC reactivities relative to isoprene is smaller. For a given (MVK+MAC)/isoprene ratio of 0.44 the contributions of the modelled OVOC reactivity ratios relative to isoprene are 0.21, 0.25, 0.27 and 0.30 for the Paulot mechanism, the kinetic reaction model, the Mozart mechanism and the RACM mechanism respectively. The four schemes would then all lie within 4%. The reactivity ratio calculated from the sequential reaction model without inclusion of fast OVOC production however would be 32% smaller (e.g., 0.17) than with fast OVOC production (0.25). The total contribution of OVOCs from the RACM model (e.g., without including the lumped hydrocarbon species OLT) is 0.20 and would be 5 to 35% smaller compared to the other three mechanisms. When constraining by photochemical age only, the RACM mechanism tends to underpredict the overall reactivity ratio relative to the other three schemes. As an example the overall reactivity ratio based on RACM would be 20%–30% lower compared to the Mozart mechanism for photochemical ages between 5e5 and 1e6 molecules/cm$^3$ days. Based on this analysis it could be as much as 30–50% lower than the OVOC reactivity from the three other mechanisms if solely based on an estimation of photochemical age.

The presented analysis suggests that a large fraction of the missing VOCs inferred from OH reactivity measurements in forested environments (e.g., 50%) could be an overestimate associated with isoprene oxidation products not accounted for by lumped photo-chemical oxidation schemes. Given the fact that many OVOCs were typically not measured during previously reported discrepancies between modelled and measured OH reactivity (di Carlo et al., 2004; Sinha et al., 2008; Mao et al., 2009) the actual missing OH sink could be much smaller and perhaps reconciled without the need to invoke large missing sources of unknown primary biogenic VOCs. For modelled OH reactivities constrained by measured OVOC/isoprene ratios the relative variation between different photochemical oxidation schemes is much smaller (e.g., 4% for the four oxidation schemes investigated here). The slower evolution of OVOC reactivity in the RACM mechanism would then largely be compensated.

4 Conclusion and summary

Our field observations in the wet season of the central Amazon Basin show that the distribution of MVK and MAC produced from isoprene oxidation is largely consistent with our current understanding of their photochemical production routes. Three explicit schemes for MVK and MAC are able to reproduce these species reasonably well. However the relative ratios among hydroxyacetone, MVK and MAC suggest that the production of hydroxyacetone is faster than current chemical photo-oxidation mechanisms can explain. The discrepancy between observations and traditional chemical isoprene degradation schemes could be reconciled by introducing direct production from isoprene and very fast secondary production as has recently been proposed theoretically.

A newly proposed isoprene mechanism that recycles HO$_x$ is largely inconsistent with observed OVOC ratios unless the thermodynamic stability of certain isoprene peroxyradicals was larger than predicted by quantum mechanical calculations.

Our measurements demonstrate the importance of accurate representation of OVOCs, which act as an OH sink and precursors for secondary organic aerosols (SOA). To date only a few studies (Spaulding et al., 2003; Warneke et al., 2001) have reported field measurements of selected OVOCs produced from isoprene chemistry. A more extensive set of OVOC observations will be necessary to fully constrain OH reactivity measurements in the future and will help improve uncertainties associated with modelling OH lifetimes in isoprene dominated environments.

Appendix A

A sequential reaction model was used to calculate the evolution of prominent isoprene photo oxidation products. Table 1 lists compounds included in the present analysis. Due to poor understanding of atmospheric oxidation of some OVOCs the sequential reaction model was restricted to first, second and third generation products exhibiting OH reaction rate coefficients greater than 1e$^{-3}$ cm$^3$/s. The selection was also based on the fact that the present analysis is restricted to photochemical ages of <2e6 molecules/cm$^3$ days, typically encountered close to isoprene emission sources. In addition the evolution of OVOCs was calculated using two lumped chemistry schemes (Mozart Chem version 4, Pfister et al., 2009; and RACM, Stockwell et al., 1997). The Mozart Chemistry scheme is used in a global atmospheric chemistry model and includes nearly explicit representation of the first and second generation oxidation products from isoprene. The Regional Atmospheric Chemistry Mechanism (RACM) was originally developed for regional air quality purposes and exhibits a higher degree of lumping for isoprene products than the Mozart Chemistry scheme.
Table A1. VOCs considered for the sequential reaction model.

| VOC                          | Rate Coefficient (295 K) [cm³/s] | Photolysis [1/s] |
|------------------------------|----------------------------------|-----------------|
| Isoprene (C₅H₈)             | 1.0E-10ᵃ,ᵇ                        | –               |
| Methylvinylketone (MVK)      | 1.9E-11ᵃ,ᵇ                       | 1.2E-6          |
| Methacrolein (MAC)           | 3.4E-11ᵃ,ᵇ                       | 2.2E-6          |
| Hydroxyacetone (HYAC)        | 6.0E-12ᶜ                         | <1E-6           |
| Glycolaldehyde (GLYALD)      | 1.1E-1¹ᵈ                          | 4E-6            |
| Methylglyoxal (MeGLYO)       | 1.7E-1¹ᵃ                         | 1E-4            |
| Glyoxal (GLYO)               | 1.2E-1¹ᵃ                         | 6E-5            |
| Formaldehyde (HCHO)          | 1.0E-1¹ᵇ                         | 4E-5ᵇ           |
| 3-Methylfuran (MF)           | 9.3E-1¹ᵇ                         | –               |
| C₅H₈O₂ (e.g. 4-Hydroxy-2-methyl-2enal) (HC5) | 5.0E-1¹ᵇ | < 1E-6 |
| Methylketone (MEK)           | 1.3E-1²ˡ                         | 7E-7            |
| Acetaldehyde (CH₃CHO)        | 1.4E-1¹ˡ                         | 5E-6            |

ᵃ Carter and Atkinson (1996)
ᵇ Gieczak et al. (1997)
ᶜ Dillon et al. (2006) (3 times larger rate constant than in Atkinson et al., 1997)
ᵈ Bacher et al. (2001)
ᵉ estimated based on Kwok et al. (1996)
ᶠ Atkinson et al. (1997)
ᵍ from Tropospheric Ultraviolet and Visible (TUV) Radiation Model: http://cprm.acd.ucar.edu/Models/TUV/
ʰ for molecular and radical branch

Table A2. Simplified sequential reaction model used to calculate OH reactivity based on VOCs listed in Table 1 assuming sufficiently high NO concentrations. Production of CO is not included. Mechanism is based on the Mozart mechanism v4 (Emmons, 2009) including additions proposed by Paulot et al. (2009b).

| Reaction: VOC+OH              | Products                                      |
|------------------------------|-----------------------------------------------|
| Isoprene (C₅H₈)             | 0.63 HCHO+0.32 MVK+0.23 MAC+Yᵢsoa HYAC+Yᵢsoa GLYALD+Yᵢsoa MeGLYO+Yᵢsoa GLYO+Yᵢso Yᵢso MF+Yᵢso HC5¹ |
| Methylvinylketone (MVK)      | 0.65 GLYALD+0.3 MeGLYO+0.3 HCHO               |
| Methacrolein (MAC)           | 0.42 HYAC+0.08 MeGLYO+0.50 HCHO               |
| Hydroxyacetone (HYAC)        | 0.91 MeGLYO+(0.125 Acetic Acid)²             |
| Glycolaldehyde (GLYALD)      | 0.2 GLYO+0.8 HCHO                             |
| C₅H₈O₂ (e.g. 4-Hydroxy-2-methyl-2enal) (HC5) | 0.5 MEK+0.25 GLYALD+0.25 MeGLYO             |
| Methylethylketone            | CH₃CHO                                        |

¹ Yᵢso=0.08±0.022: Yield based on estimated yields for hydroxyacetone. Yᵢso=0.45×Yᵢsoa;
² Further photooxidation of acetic acid is neglected due to slow reaction rate.
Table A3. Initial radical and peroxyradical reactions according to Peeters et al. (2009). The asterisk (*) symbol indicates an R or RO₂ radical. Further reactions of MVK, MAC and hydroxyacetone are treated as a first order reaction with respect to OH according to Table 2. Only the OH-addition to carbons 1 and 4 in isoprene are considered here corresponding to 90% of oxidized carbon. This is reflected by the radical yields after initial OH addition. Reaction constants and rates used for the calculations are listed in the last column. The reaction rates kf and kr represent the forward and reverse reactions for O₂ addition and elimination and the forward reactions for the 1,6 and 1,5-H-shifts.

| Reaction VOC | Products | k_{OH} (cm³/s) |
|--------------|----------|----------------|
| Isoprene + OH | 0.3 t-1-OH-isoprene*+0.3 c-1-OH-isoprene*+0.21 c-4-OH-isoprene* | 1.0E-10ᵃ |
| O₂ addition/elimination | kf (10^{-12})/kr (cm³/s)/(s^{-1})ᵇ |
| t-1-OH-isoprene*+O₂ | E-1-OH-4-OO*+isoprene | 0.3/15.0 |
| c-1-OH-isoprene*+O₂ | 1-OH-2-OO*+isoprene | 1.5/7.0 |
| t-4-OH-isoprene*+O₂ | E-4-OH-1-OO*+isoprene | 1.5/30.0 |
| c-4-OH-isoprene*+O₂ | 4-OH-3-OO*+isoprene | 1.5/1.9 |
| RO₂+NO | k_{NO} (cm³/s)ᶜ |
| E-1-OH-4-OO*+isoprene | products | 8.0E-12 |
| 1-OH-2-OO*+isoprene | 0.93 MVK | 8.0E-12 |
| Z-1-OH-4-OO*+isoprene | products | 8.0E-12 |
| E-4-OH-1-OO*+isoprene | Products | 8.0E-12 |
| 4-OH-3-OO*+isoprene | 0.93 MAC | 8.0E-12 |
| Z-4-OH-1-OO*+isoprene | 0.5 hydroxyacetone | 8.0E-12 |
| RO₂+HO₂ | products | 8.0E-12 |
| RO₂+RO₂ | products | 8.0E-12 |
| RO₂ H-shifts | kf (s^{-1})ᵇ |
| 1-OH-2-OO*+isoprene (1.5-H-shift) | OH+HCHO+MVK | 4.0E-03 |
| 4-OH-3-OO*+isoprene (1.5-H-shift) | OH+HCHO+MAC | 1.1E-02 |
| Z-1-OH-4-OO*+isoprene (1.6-H-shift) | HOO + hydroperoxymethylbutenal | 1.0 |
| Z-4-OH-1-OO*+isoprene (1.6-H-shift) | HOO + hydroperoxymethylbutenal | 8.0 |

ᵃ Carter and Atkinson (1996); Gieczak et al. (1997)
ᵇ Peeters et al. (2009)
ᶜ Emmons et al. (2009)
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