In this review, we examine house dust and residential soil and their use for identifying sources and the quantifying levels of toxicants for the estimation of exposure. We answer critical questions that focus on the selection of samples or sampling strategies for collection and discuss areas of uncertainty and gaps in knowledge. We discuss the evolution of dust sampling with a special emphasis on work conducted after the publication of the 1992 review by McArthur [Appl Occup Environ Hyg 7(9):599–606 (1992)]. The approaches to sampling dust examined include surface wipe sampling, vacuum sampling, and other sampling approaches, including attic sampling. The metrics of presentation of results for toxicants in dust surface loading (micrograms per square centimeter) or surface concentration (micrograms per gram) are discussed. We evaluate these metrics in terms of how the information can be used in source characterization and in exposure characterization. We discuss the types of companion information on source use and household or personal activity patterns required to assess the significance of the dust exposure. The status and needs for wipe samplers, surface samplers, and vacuum samplers are summarized with some discussion on the strengths and weaknesses of each type of sampler. We also discuss needs for research and development and the current status of standardization. Case studies are provided to illustrate the use of house dust and residential soil in source characterization, forensic analyses, or human exposure assessment. 

**Key Words:** chemical composition, exposure characterization, house dust, physical composition, residence wipe sampling, source identification, vacuum sampling. 

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The maturation of the field of exposure analysis and assessment has led to the development of many innovative methods and strategies to detect, reduce, and prevent human contact with environmental toxicants. These toxicants can be found in air, water, soil, or food. Some methods employed by investigators are new (e.g., biological markers and human videotaping), whereas others are adaptations of methods used for general or specific types of environmental or occupational investigations (microscopic analyses, modeling, and personal monitoring). Depending on the type of method chosen, each may be applicable for epidemiologic investigations, human exposure profiling for specific contaminants or events, and/or risk assessment. In addition, and as a result of the findings made in the Total Exposure Assessment Methodology (TEAM) studies and others conducted during the 1980s, the field of exposure analysis has turned its attention beyond primarily measuring toxicants in outdoor environmental situations (1–9). Now information is frequently obtained from one or more different microenvironments occupied or passed through by an individual or individuals over the course of a day.

The purpose of this article is to review the types of information available on dust in residential settings, the metrics used for exposure and source characterizations, and the composition of dust. We also review and evaluate the methods used to affect residential dust and soil. We explore the applications of dust for exposure and source-related analyses and ways to improve quantitative dust characterization in the future.

A number of studies have shown that for individuals and populations, especially children and other vulnerable subgroups, the home environment can be a potential source of passive or active exposure to toxicants (10–23). This presents a complex situation for the exposure specialist and for other environmental health scientists in their attempts to identify the potential or actual risk associated with common or specific environmental toxicants. There are many different types of toxicants that can be present for short or long periods in a building and each could be in a form or location that is readily accessible for contact by inhabitants or visitors (5,24–26). Because toxicants can come from multiple sources (indoors or outdoors), media, and activities, exposures can occur at single or multiple routes of entry into the body. The question that arises is, how does one determine which toxicant and source, or multiples of each, are of greatest concern to the health of the residents?

Beyond obtaining measurements for the toxicant levels present at routes of entry into the body, variables needed to characterize exposure are the activities and the patterns of activity and uncommon events that may result in passive or active exposures. Thus, the number and complexity of the types of variables needed to understand exposure can become large. This leads directly to consideration of a triage to the application of sampling and analysis strategies for finding the best ways to identify, examine, and explain potential or actual exposures and to find and select the most appropriate mitigation or prevention strategies. Sampling and analysis strategies are becoming increasingly important as the field attempts to provide information on cumulative exposure to multiple toxicants in a single medium and the aggregate exposure associated with a single toxicant from multiple exposure pathways (26–29).

Because of the multiplicity of indoor and outdoor sources and routes of contact that can occur in the home, residential dust (house dust) and residential/community soil represent two categories within one medium—soil—that can benefit from a triage of measurement strategies. The approach will allow the investigation of the influence of multiple variables on the intensity and duration of contact with toxicants that can result in cumulative or aggregate exposures.

**Soil and Dust: Perspective for Exposure Studies**

For many years, soil samples have been collected as cores in various urban and industrial locations (30). There is a wealth of information on work conducted after the publication of the 1992 review by McArthur [Appl Occup Environ Hyg 7(9):599–606 (1992)]. Now information is frequently obtained from one or more different microenvironments occupied or passed through by an individual or individuals over the course of a day.

The purpose of this article is to review the types of information available on dust in residential settings, the metrics used for exposure and source characterizations, and the composition of dust. We also review and evaluate the methods used to affect residential dust and soil. We explore the applications of dust for exposure and source-related analyses and ways to improve quantitative dust characterization in the future.

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on the chemistry, physics, and microbiology of soils (31–33). The analytic results obtained from many soil samples have been used to determine the levels of various materials and toxicants at a certain depth (0 to > 100 ft) from the surface, but most of this work does not deal directly with exposure characterization. Opportunities for using this type of data in exposure assessment include quantifying significant penetration of toxicants into the groundwater that is used for a drinking water supply or the diffusion of volatile material into a basement. Most of the measurements made in soil at a distance beyond 1 ft below the surface are of little value in assessing direct dermal exposure (34,35). Data or soil contamination, however, is still important for profiling the distribution of material to various depths, identifying sources, and assisting in defining the period of time when the deposition and/or accumulation of toxicants occurred in the soil. Core sampling is also used to detect the presence of a toxicant or tracer material in locations that include ice or snow packs and sediments below bodies of water (26). Numerous soil sampling and analysis programs have been conducted at hazardous waste sites (e.g., Superfund sites) and abandoned industrial sites (e.g., brown fields). These programs have been designed to establish the horizontal and vertical extent of contamination before remediation (36). In addition, research studies have been completed in arid regions around the world. These have focused on the reentrainment of desert sand and its redistribution to locations > 1,000 km away (37–39). However, the focus here is on the ability of sand to be used to assess higher exposure, with the primary purpose of accounting for the soil component of residential dust.

The results of soil characterizations at hazardous waste or urban sites have been used, and in many circumstances continue to be used, as the basis for exposure characterization in the preliminary risk assessments required before site cleanup (26,40–42). However, in addition to questions about the appropriateness of these data for a specific exposure assessment, the analytic plan usually employs default factors when estimating intensity and duration of contact with individual or multiple toxicants. The results of such analyses usually provide very crude, and in some cases unreasonable, worst-case estimates of possible exposures and risks (43). Over the past half decade, however, efforts have been made to encourage the use of more site-specific data to determine plausible patterns of activity that lead to exposure to the toxicants that are present in soils (44).

By collecting samples from the surface of a yard, a park, open or abandoned spaces, or locations downwind of an explosion or fire, characterization of toxicant levels in the soils have improved human exposure assessments. One use of dust collection and analysis approaches has been to characterize the dust emitted in downtown Manhattan from the aftermath of the World Trade Center attack (45).

Initially, measurements in surface soil samples were used to detect the levels of a toxicant that were available for resuspension into the atmosphere and then used to estimate inhalation exposure (46–48). Now such measurements are also applied to estimate direct dermal contact and to determine the potential ingestion or skin absorption of a toxicant found in or on dust or soil as a chemical residue (24). In many circumstances, the levels of chemical, physical, or biological contaminants in or on surface soil will represent the material that has been recently deposited on the soil and the fractions of soil that can be most easily transferred indoors. Further, work by Wallace et al. (2) during the TEAM study; by Charnay et al. (49) and Roberts et al. (50) on lead exposure; by Simcox et al. (16) for pesticide exposures on farms; by Freeman et al. (19,51) on chromium exposure; and by Pellizzari et al. (52) and Bonanno et al. (53) for multiple pollutant and multiple media exposures during the National Human Exposure Assessment Survey (NHEXAS), to name a few, have demonstrated that a complex series of interactions takes place between home characteristics and activities, which can alter toxicant exposure, or indoor and outdoor source strengths.

During the early part of the twentieth century, a convenient method for collecting total particulate matter (undifferentiated by particle size) in air that had deposited on surfaces was the dust bucket. It was a simple device, just an open water bucket placed on a roof or other secure location to collect material available in the air for deposition on a surface. The samples were usually analyzed by visual observations, color, and total content, and then subjected to simple chemical analyses. A bucket would remain out for at least a month to collect all the material (including bugs, etc.) that deposited from the atmosphere via wet and dry deposition (54). The results were semiquantitative at best and usually difficult to interpret. Fortunately, devices used to collect dry and wet deposition have improved over time, and some are designed to make quantitative determinations of outdoor atmospheric deposition. These modern devices are currently used in national and international monitoring networks sponsored by various agencies (54). Indoors, deposition plates and other passive collection devices have been used to quantify the contribution of outdoor and indoor sources of particulate matter onto indoor surfaces (55).

Once researchers began to consider human exposure in residential neighborhoods and on farms, field studies were designed that included potential or actual soil contamination, and they focused on quantifying the movement of resuspended, contaminated outdoor soil and road dust into indoor locations (46–48,55–61). These studies were designed to examine the exposure and risk posed by toxicants deposited in the lungs of people living in zones potentially affected by point and area sources. Eventually, hypotheses about the transport of dust and dirt indoors, and contact by residents, led to parallel efforts on developing and employing techniques to measure the levels of contaminants in house dust after it has been generated or transported indoors and settled on surfaces or has been tracked into the home (15,16,23,52,53,62–71). In urban areas, some field studies have characterized the movement and accumulation of lead and other metals indoors (10,19,71–76). Based on the results of these studies, important transport processes to consider include the resuspension and transport of toxicants indoors, including tracking toxicants present in the street or in yard dust or soil by residents or pets (25,26).

For farms and residential lawn applications, some investigations have focused on the drift of pesticides to homes after spraying and on the tracking and deposition of resuspended outdoor dust or surface soil indoors (23,66). Recent exposure assessments for lead, pesticides, and polycyclic aromatic hydrocarbons (PAHs) and other toxicants (e.g., asbestos) also include characterization of exposure that resulted from toxicants deposited in and on objects found in homes. As a result, house dust is being used as a tool to assess environmental health risks by providing answers to questions about the potential for personal contacts that may have led to a significant exposure for one or more routes of entry into the body. Further, such information is now being used to examine the liability issues related to residential toxicant exposure and will eventually lead to development of standardized protocols for use in quantified forensic exposure characterizations.

**General Composition of House Dust**

When residential dust is selected as a metric of exposure, questions include:

- What are the components of residential dust?
- What are the behaviors and activities of the individuals that live in a particular location that either led to exposure or resuspension of dust?

If there was a homogeneous distribution of material in a home, and a major fraction of the dust was similar in composition, it would be easier to conduct analyses of exposure. At this time, however, limited data are available that can accurately describe the distribution of basic major constituents of house dust in American residences or in residences in other countries.
Depending on the situation and the types of indoor/outdoor or personal sources associated with a residence, the toxicants present in the dust can include semivolatile and non-volatile pesticides, PAHs, heavy metals, persistent organic compounds, asbestos, and viable biological particles (16, 18, 22, 64, 65, 68, 80). The sources of these compounds include combustion, professional product applications, typical residential product uses, fugitive emissions, transport, and degradation of plants and animal and insect parts. In fact, the elemental nature or persistence of some of these materials allows an investigator to compile a historical record of material deposited in the carpet. Clearly, in some cases this could reveal significant loadings caused by indoor or outdoor sources, but a historical analysis could also provide fingerprints for specific toxicants, or general internal or external sources of toxicants. When investigating a home that has environmental problems or when attempting to characterize general patterns of exposure, baseline data will also improve forensic-based sampling for source identification or exposure characterization.

### Historical Perspective on Dust Sampling

The original techniques used for indoor dust collection were developed before 1970, and the majority were wipe-sampling techniques. Until recently, however, none of the techniques had undergone much scrutiny or rigorous validation. A major review of dust contaminant sampling techniques, primarily wipe-sampling techniques, was published in 1992 by McArthur (81). His evaluation identified wipe sampling for detection of radionuclides in the laboratory and other nuclear facilities as one scenario in which dust sampling could be used to routinely monitor health and safety.

### Table 1. Examples of typical nonexperimental studies or postremediation levels of dust and selected toxicants found indoors: loadings and concentrations.

| Compound                      | Method                        | Concentration range (Loading) |
|-------------------------------|-------------------------------|-------------------------------|
| Dust                          | HVS3                          | 0.32–14.4 g/m²                |
| Adgate et al. (17)             | LWW wipe: floor               | 0.05–7.0 g/m²                 |
| Roberts et al. (74)            | HVS3: rug/typical home vacuum | <1.0–26 g/m²                  |
| Farfel et al. (87)             | HVS3: floor                   | 0.01–90 mg/m²                 |
| Roberts et al. (50)            | HUD-wipe: floor               | 0.09–60 mg/m²                 |
| Roberts et al. (50)            | HUD-wipe: windowsill          | 0.05–600 mg/m²                |
| Roberts et al. (50)            | LWW wipe (windowsill)         | 0.01–45,000 mg/m²             |
| Adgate et al. (17)             | Vacuum                        | 75–700 µg/g                   |
| Pesticides                    |                               |                               |
| Rudel et al. (80)              | Carbaryl                      | 1.26–89 µg/g                  |
|                    | O-Phenyl-phenol               | 27.2–140 µg/g                 |
| Liow et al. (79)               | LWW wipe: floor               | 0.01–4.18 µg/m²               |
| Chlorpyrifos                  | EL sampler: carpet            | 0.02–44.5 µg/m²               |
|                               | EL sampler: floor             | 0.03–36.6 µg/m²               |
| PAHs                          | HVS3                          |                               |
| Rudel et al. (80)              | Chrysene                      | 0.455–10.6 µg/g               |
| 12 PAHs                       | HVS3                          | 2–12 µg/g                     |
| Dust mites                    | HVS3                          | <0.2–0.94 µg/g                |
| Fungi and microorganisms      |                               | 0.11–3.46 µg/m²               |
| Molhave et al. (78)           | Molds                         |                               |
| Total microorganisms          | Vacuum bags                   | 130,000–160,000 CFU/g         |
| Fungi                         | Vacuum bags                   | 71,000–90,000 CFU/g           |
| Phenols                       |                               |                               |
| Rudel et al. (80)              | Bisphenol A                   | 0.25–0.48 µg/g                |
|                               | Phthalates                    | 0.17–6.82 µg/g                |
| Rudel et al. (80)              | Di-2-ethylhexyl phthalate     | 69.4–524.0 µg/g               |

Abbreviations: EL, Edwards and Lioy; HUD, Housing and Urban Development; HVS3, high-volume surface sampler; LWW, Lioy-Weisel-Wainman.
However, qualitative wipe samples were used as early as 1940 in hospitals to determine microbial levels in operating rooms and to evaluate cleanliness of pharmaceutical manufacturing facilities (26,67,82). The first technique used for wet wipe sampling of contaminants present on dust-laden surfaces was published by Vostal et al. in 1974 (83).

The conclusions made by McArthur (81) in his review were enlightening because he stated that surface sampling has limited reliability for use in exposure assessment. Further, there was an important final quote in his paper that was attributed to H. J. Dunster, who nearly 40 years earlier recommended that hygienists pause and consider finding a better way (84):

Surface contamination monitoring has not become a tool of occupational hygiene in general, partly because of the difficulty of monitoring for non-radioactive materials on surfaces. If the techniques of measurement were easier it is likely that the occupational hygienist would find monitoring of surface contamination to be a technique as useful to him as it is to his radiological colleague.

Unfortunately, there was not enough interest during the 1960s to follow through on Dunster’s observation. Based on MacArthur’s analyses and review of the standard approaches used in the field during the mid-to-late twentieth century, it was clear that there was an inability to employ reliable surface samplers for quantitative measurements of occupational or community exposure. Fortunately, during the late 1980s and early 1990s, work was completed by a few investigators that improved the methods for surface sampling: new techniques for surface sampling and rug sampling (Roberts et al. (15), Lioy et al. (85)); the size-selective rug vacuum sampler [Lewis et al. (86)], the EL press sampler [Edwards and Lioy (87)], and a number of chemical-specific samplers [e.g., for pesticides (23,74)].

### Table 2. Potential carcinogenic, neurotoxic, or endocrine-disrupting compounds in fine carpet dust.

| Compound       | Dust concentration (ppm) | Carpet loading (µg/m²) |
|----------------|--------------------------|------------------------|
|                | No. of homes | Detected (%) | Median | Max | Median | Max |
| Metals         |             |              |        |     |        |     |
| Arsenic        | 25          | 100          | 5.15   | 16.70 | 5.11   | 91.90 |
| Cadmium        | 24          | 100          | 6.75   | 20.40 | 9.10   | 113.00|
| Mercury        | 24          | 100          | 1.69   | 15.90 | 2.87   | 44.7 |
| Lead           | 25          | 100          | 164    | 1,200 | 202    | 6,120 |
| Copper         | 25          | 100          | 170    | 1,144 | 260    | 1,790|
| Sum of PCB congeners in house dust | 32       | 100          | 0.42   | 3.60 | 0.26   | 17.00 |
| Sum of seven probable (B-2) human carcinogens in house dust | 10       | 100          | > 200  | > 200 | NA     | NA   |

**Note:** Formulations from various studies reviewed by Roberts (74).

### Table 3. Summary data on composition of house dust in seven U.S. cities collected by high-efficiency vacuum cleaner.

| Characteristic | Result          |
|----------------|-----------------|
| Range (% by gravimetric analysis) of fibrous particles | 9–89 |
| Range (% by gravimetric analysis) of non-fibrous particles | 11–91 |
| Size range (% nonfibrous of total particles collected) > 300 µm | 4–83 |
| 75–300 µm | 1–32 |
| < 75 µm | 0–20 |
| Days since last cleaning (average) | 14.2 |
| Days since last cleaning (range) | 1–150 |
| No. of people living in home (average) | 3.3 |
| No. of people living in home (range) | 1–10 |
| Composition (qualitative) by polarized light microscopy | Skin, soil, starch, hair, cotton, plant (> 85% of samples) |
| Most frequently identified materials | Fungal material, synthetic fibers, polymers, paint, metals |
| Second most frequently identified materials | Cities were San Diego, CA; Columbus, OH; Phoenix, AZ; Miami, FL; New York metropolitan area (including New Jersey); Denver, CO; Kansas City, KS. Included were 36 individual home samples and 12 sets of pooled samples. Samples were collected in the kitchen, living room, and/or bedroom, or another room other than the kitchen (79).
assessment and/or residential source identification can employ one or both of the following design options: a) a device that mimics the ability of the hand to pick up and retain contaminants on the particles that are found on many surfaces, and b) a device that will quantitatively collect all or a reproducible fraction of contaminants found on specific surfaces.

Currently, it is impossible to select one wipe sampler that functions properly under all conditions. For example, a sampler that can be used exclusively for smooth-surface sampling would not necessarily have the same design as a sampler used for textured surface sampling. It is straightforward to find a sampler that will retain the dust, dirt, and other particles present on a smooth surface. For textured surfaces, one has the added problem of ensuring that contact is made between the sampler and the entire surface, including the crevices. In both instances, a smooth or textured surface, an investigator must also try to mimic the levels of contaminants picked up by the hand. Thus, the sampler needs to be designed with surface characteristics that can collect the particle size distribution that best represents retention by the hand, after wet and/or dry contact with an object or floor.

Potential Exposure Variables or Metrics Measured for Residential Dust and Soil

Applications of household dust in exposure characterizations and assessments require a sampler that can measure one of two metrics (or variables) and preferably both. The first is the loading of materials on a surface in units of micrograms of material or toxicant per square centimeter of surface. To date, the metric “surface loading” has received the most attention during the development of techniques because it is the simplest quantity to measure. At a minimum, it tells the investigator whether a toxicant is present at a location, and it can provide values in micrograms per square centimeter for the distribution of a toxicant in a room on surfaces or throughout the entire residence. The second quantity is the “concentration of a material in the collected dust” collected by the sampler. It is reported as micrograms of a toxicant per gram of dust. This value is much more difficult to quantify because the collection medium must be pre- and postweighed under standard weighing conditions, and the medium must not change character during dust collection.

Both surface loading and dust concentration are valuable metrics in characterizing exposure. Measurements of surface loading can be used to estimate the amount of material available on a surface and the amount available for contact by a person. Dust concentration can be used to characterize potential sources and source types inside or outside a home. In each case, there are limitations and uncertainties based on the design and validation of each device for the intended purpose of sampling; the information available to characterize activities and activity patterns and actual source use; and the area on a surface in a room or in the residence that the samples actually represent. Some typical data collected on dust loading and the concentrations of common toxicants found in house dust are previously shown in Tables 1 and 2.

Sampling Methods

Surface wipe sampling. Almost any surface found in or around a residence can come into contact with some part of the human body, and each would be a candidate for surface wipe sampling. Thus, the investigator must clearly articulate the purpose of the sampling study. These will guide selection of the appropriate surfaces to sample, the toxicants to be measured, and the frequency of sampling. For example, if one were examining a child’s contact with a toxicant, the focus would be on objects and surfaces that are frequently used, touched, or mouthed by the child during the day (88).

A number of devices have been used to examine the levels of toxicants on surfaces. Depending on cleaning frequency, the materials present on surfaces to be sampled can reflect deposits caused by many different activities that occurred over a range of time periods. Undisturbed surfaces (e.g., top of a refrigerator), can be indicative of materials deposited over a long period of time, whereas frequently cleaned surfaces (e.g., kitchen countertops) will be indicative of the most recent deposits. Toys and play surfaces or work surfaces can be indicator surfaces for objects most frequently contacted by children and adults, respectively. Windowsills and wells can provide information on materials that are carried from outdoors to indoors or may reflect the materials in paint flaking from the surfaces around the window. Thus, surface sampling can provide a wealth of material for estimating potential contact with any levels of toxicants of concern for acute or chronic health effects, aggregate or cumulative exposure assessments, or identification of sources. The ability to identify sources is sometimes difficult using wipe samples because of the relatively low levels of dust usually found on many surfaces.

Rug and carpet sampling. When the analysis of exposure moves beyond surface sampling of floors, tables, and so on, an investigation can focus on toxicant levels in rugs and carpets and other plush objects such as pillows. These residential furnishings pose another set of challenges for the researcher to address, both conceptually and analytically. For example, carpets and rugs have major structural features that require the investigator to first establish what the material deposited in a carpet actually represents (74). The first feature of a rug or carpet associated with dust retention is its surface and the easily accessible layers of carpet fibers. The second is the base that binds the carpet or rug fibers together. From the standpoint of variables used for research or forensic application of exposure analysis and assessment, each provides variables and results that can be interpreted in different ways.

When the material sampled from the surface layers of the rug is analyzed for dust mass and composition, it is usually done to determine the levels and types of material that are accessible or available for contact with a human hand or other parts of the body. The results can also be used to represent the layers of a rug that can contact edible materials (e.g., food) that are rubbed or dropped on the rug by a potentially exposed person and removed by friction (89). In some instances, materials present on a rug are called dislodgable dust, but the definition does not truly describe all contact issues.

Resuspendable dust can also be deposited onto clothing or skin or scavenged by food and other mouthable objects that contact the rug surface (90). Therefore, any material or toxicant attached to the surface of a rug or carpet can be collected and quantified. Subsequently, it may be appropriate to use the results as an indicator of a potential dermal contact and nondietary intake by adults or children. The latter requires the collection of information on activity patterns and the intensity and frequency of contact with a toxicant to adequately assess a person or population’s exposure.

The major challenge for exposure analyses is ensuring that the sampler retrieves material that is indicative of the types and levels that a person comes into contact with on the rug. For example, the type of rug (e.g., loop level or shag) will have different retention characteristics for particles on the rug surface. The surface can also significantly affect the efficiency of removal of particles from the surface by a vacuum cleaner (74). The problems posed by such situations include ensuring that a true indication of the levels and distribution of material deposited on the surface of rug fibers is obtained during sampling and that the sample of dust is representative of material available for removal. Finally, work by Wang et al. (91) has shown that the collection efficiency of a vacuum cleaner is affected by the relative humidity in the home.

In contrast to the surface of a rug or carpet, the material embedded at the base of a rug or carpet can provide information needed to complete exposure analyses. The embedded material can be considered to be equivalent to the weathered loading of soil and air pollution deposited outdoors on surfaces over
time via atmospheric deposition at the base of a forest canopy or within the sediments of a lake. In this case, the loading of particles at the base of a rug would be affected by particle deposition, dynamic removal and redeposition of material, and spills.

An individual rug does not provide a complete historical record of the loading that could theoretically accumulate over time. However, a rug/carpet can provide a record of indoor and outdoor sources that have contributed to the overall levels of dust and components of dust that were tracked into a residence, were spilled or applied to the surface, or were removed from the air over time. Rugs and carpets also are major reservoirs of material because their large total surface area increases the total amount of material deposited on or stored in the fibers. Potentially rugs and carpets have levels of specific materials that are available for resuspension or reemission (semivolatile compounds). Thus, rugs or carpets can be used as a research or a forensic tool.  

The most important variables to consider before attempting to sample a rug or carpet are type (e.g., level loop or shag), age, condition, location in the residence, cleaning frequency, and family history. An individual rug could provide a variety of long-term or short-term data and information on the types of activities and sources that affect toxicant levels in an individual home.

Similar to outdoor situations, materials present in the rug or carpet can be resuspended from the rug or carpet. However, the mechanisms of resuspension are different from those associated with outdoor air. For example, the reentrainment of carpet dust, both surface and embedded, occurs using the typical household vacuum cleaner. Vacuum cleaners that do not use a HEPA filter system and do not have a sealed capture system will not efficiently pick up and retain the fine particles in carpet dirt and dust (92). These include most devices generally used by the public. Some fine particles will be resuspended and then redeposited on the floor and other surfaces. Thus, over time any residuals from major spills or applications that remain after clean up or degradation will eventually be reduced in magnitude but will also contribute some amount of mass to the general long-term material burden in the rug. The last point also brings up the consideration that if one is attempting to look at outdoor influences on the rug burden, it is important to include sampling locations near high-traffic doorways. This was recently done by Bonanno (93) and by Farfel et al. (94) for lead, cadmium, and arsenic.

Work done at the Environmental and Occupational Health Sciences Institute (EOHSI) and the U.S. Environmental Protection Agency (EPA) has also shown that semivolatile materials initially deposited on a floor or in a crack or crevice will evaporate and condense from the point of application onto other surfaces (64,66,95). This is of particular concern for pesticides sprayed in the home. Over time a lower concentration of the semivolatile compound will be spread to a variety of surfaces in a room or residence that were not initially sprayed with pesticides, thus increasing the possibility of exposure to residents and visitors.

In a recent study by researchers at the Harvard School of Public Health (80), a large number of chemicals were measured and the viable biological particles in vacuum cleaner samples obtained in residences were tested. The investigators also measured the levels of specific toxicants in the air. The objective of the study was to quantify the levels of hormonally active agents and mammary carcinogens in homes. A similar, but more focused, hypothesis-driven investigation was conducted for microbiologicals and fungi obtained from rugs in offices located in Denmark (78).

Other sampling approaches. A number of techniques in addition to wipe samplers and vacuum samplers can provide data for research, regulatory, risk reduction, and forensic analyses of population and individual exposures. Each can be used to collect dust from undisturbed locations within a residence or associated structures (e.g., garages, utility sheds). One of the best examples, but one that has been rarely used in exposure analyses, is attic dust. When one considers that the normal life of a residential roof is 20–30 years or longer, if you add multiple layers of roofing materials, the attic can contain a record of undisturbed archived deposited particles. Particles would have infiltrated the residence by diffusion and advection through the eyeways or other passive ventilation portals (indoor or outdoor), and then settled on surfaces. Attic dust has been used periodically to examine the deposition of radionucleides from nuclear fallout and nuclear power plant emissions. Recently, studies have examined attic dust for the levels of toxicants that may have been emitted to the ambient air, transported, and then deposited in the attic (96,97). Other studies examined deposition in attics located in Nevada (98) or downwind of the Chernobyl nuclear accident (59,99–101). Each included measurements of radionucleides emitted by nuclear arms tests and/or nuclear power plants, metals emitted from point or area sources, and specific tracers emitted by from industries.

Status and Issues for House and Building Dust Samplers and Their Use

There are a variety of samplers available to collect indoor-generated or outdoor-generated house dust, yard, or street dust. However, no one sampler has been invented that will collect material from all surfaces. Further, many samplers are not designed to take the same type of sample. In fact, although a number of samplers have been developed, most have never been tested for performance in terms of what the collected material is supposed to represent. Some devices have been compared during field studies to test sampler collection performance characteristics, but usually this is done only with respect to the levels of a compound or mass that other samplers generally collect. In general, the devices used in health physics, industrial hygiene, environmental hygiene, and exposure assessment measure surface loading as the amount of a toxicant per square centimeter. However, in evaluating sampler performance, it is important to evaluate and identify the method used for operating the sampler. Clearly, without performance testing for specific physical/chemical properties, it is not possible to obtain the information necessary to determine what the measured values in a sample actually represent for a surface or exposure. One also requires information about the character of the surface sampled and the atmospheric (environmental) condition at the time of sampling. Over time, a level of consistency between samplers used in a study and sampling location can be established through intercomparison studies and side-by-side sampling (102–105).

For example, a baby wipe has been used as a wet surface sampler (102). It will scour any material from the surface, and possibly the subsurface, but the results obtained will be independent of many variables that affect accumulation on the hand. This is just one example of an issue associated with all surface samplers. Such a problem, however, does not preclude using the results for forensic analyses or research in exposure analysis; it just requires definition of the applicability of the measurements or what aspects of the results increase or decrease the uncertainty of an assessment.

Wipe samplers. McArthur (81) summarized wipe sampling methods available in 1992, and a modified list is provided in Table 4 (106–119). The variety of sampler types and solvents used to collect the material for a surface is large, and some are more applicable with one toxicant, whereas others can be tailored to collect and provide measurements for a number of toxicants. Fenske et al. (14) commented that the precision of wipe sampling would be improved by defining the area and standardizing the materials and methods used for sampling.
We have developed two devices for surface dust sampling. The first was the Lioy-Weisel-Wainman (LWW) dust wipe sampler (85). It was designed and tested to quantitatively collect all materials deposited on flat surfaces, and the design eliminated the confounding influence of operator pressure on the amount of dust pickup by the sampler. The LWW was also the first wipe sampler that could quantitatively establish both the concentration (in micrograms per gram) in dust and the surface loading (in micrograms per square centimeter) of dust from the same sample. Its applications are confined to sampling horizontal or vertical flat surfaces, although the basic substrate used for the collection from the surface, a nucleopore 50 mm × 55 mm Perkin-Elmer drain disc (Perkin-Elmer, Norwalk, CT), has been tested and found to be durable enough to collect material present in a window well. The sampling substrate used by the LWW can be changed to efficiently collect specific toxicants (e.g., emeproc carbamin-18 discs; 3M, Minneapolis, MN) for the collection of pesticides (120).

The LWW also operates as a wet or dry collection substrate for sampling specific surfaces. Caution must be exercised, however, when selecting a wetted substrate because water and/or solvents can ruin finished surfaces frequently found on furniture and floors. In addition, if a surface is painted and the paint is in poor condition, a wipe can take off small chips of paint. This may give anomalously high levels of lead in dust, but the data cannot be ignored because lead paint chips present a hazard. Yen et al. (75) found that lead loading varies with time of the year. Finally, Paustenbach (26) stated that the development of the LWW provided “fairly sophisticated work to standardize (wipe sampling) procedures.” Applications of the LWW within various studies have found the loading of dust on flat surfaces to be associated with biomarkers of exposure (20) and with cleaning practices within the house (121,122).

Another issue is the protocol used for sampling dust. This was intensively examined and incorporated as a major feature in the design of the LWW sampler patent (123). The features of the sampler (no operator pressure and a defined template) were rigorously tested in the design phase of the LWW and achieved reproducibility for a standard dust of a coefficient of variation between 6 and 25% for replicate samples (86). Freeman et al. (124) demonstrated that for side-by-side field samples, the mean difference was 35%. This was in an uncontrolled situation; thus there is large potential for variability of dust loadings in adjacent locations. The initial versions of the LWW were a bit cumbersome to use, but eventually a modified design was employed to the collection of samples in approximately 300 homes during the six Midwestern State NHEXAS pilot study, and the samples were analyzed for elements and pesticides (52). In contrast, most other devices used for wipe sampling, including cloth wipes, gauze pads, and gloves, have not received the same level of performance testing.

After the success of the LWW sampler, it was important to go beyond collecting all the dust deposited on a surface to make inferences about the dermal uptake of a toxicant (88). Thus, we developed the Edwards and Lioy (EL) sampler, which had a sampling surface designed specifically to mimic the collection efficiency for the size distributions of particles on the surface of the human. It had an application pressure coincident with the pressure placed on a surface by a human. The first model of the device focused on the “nonsticky hand” (i.e., dry), and it was designed to collect a sample composed of repeated presses on a surface. It has been successfully used to determine the amount of dust and pesticides picked up from various types of textured and smooth surfaces (e.g., floors and carpets) (120).

Edwards and Lioy (87) made extensive comparisons with other wipe samplers (adhesive tapes, cotton gloves, etc.) to determine their utility as a metric of hand exposure to dust present on surfaces. The metric for calibration was the final particle size distribution retained on a human hand for a series of controlled experiments. This was the first attempt to determine what types of loadings could be transferred from the floor to a hand by a sampler to eventually provide a device that could eliminate use of the human hand as the dermal sampling device. Up until now the only effective way of completing dermal sampling was to sample the actual hands of adults and children after contact with various surfaces. The hand rinse method can have large uncertainties. For example, if one does not know the history of activities, including hand washing, and surfaces contacted prior to washing the hand, it is difficult, if not impossible, to accurately assess exposure. Further, in some cases the hand is washed with a liquid other than water, which requires added scrutiny by institutional review boards.

Edwards and Lioy (87) showed that methods typically used for wipe and surface sampling did not have the same particle size distribution as that retained by the human hand. The results in Figure 1 indicate that only synthetic skin and the C18 filter used in the EL sampler closely mimicked the uptake and retention characterization of the dry hand for particles < 250 μm in diameter. This was anticipated because none of the samplers was designed specifically to mimic exposure to house dust. The final version of the EL sampler collected a size distribution of the particles equivalent to that attached to the human hand for a “non-saliva—laden” condition of the hand. The studies to date have shown that if one wants to mimic hand retention with a surface sampler, the total burden on the floor or other surfaces is not representative of what will be found on a human hand.

When sampling a surface with the EL sampler, there was no continual and uniform buildup of mass. In fact, after about four to five presses, little new mass was added to the surface loading. The results of Rodes et al. (124) indicate that at some point new contacts with a surface just replace some of the

| Reference | Compound | Sampler type | Sample area (cm²) | Solvent |
|-----------|----------|--------------|------------------|--------|
| Alexander et al. (106) | TCDD | Glass fiber | 625 | Dry |
| Bently et al. (107) | Chlorophenol | Whatman 1 | 14.2 | Dry |
| Chalk et al. (108) | MDA | NA | — | — |
| Chavalitnikul et al. (109) | Lead | Whatman 42 | 100 | Water |
| U.S. EPA (110) | Misc | Cotton swab | 5 × 19.6 | Acetone |
| Fenske et al. (111) | Chlorophenol | Surgical gauze | 223 | Water |
| Fenske et al. (114) | Chlorpyrifos | Surgical gauze | 100 | Isopropanol |
| Hryhorczuk et al. (112) | PCBs | Whatman | 900 | Hexane |
| Lees et al. (113) | PCBs | Glass fiber | 900 | Hexane |
| Lichtscheider et al. (114) | Misc | Whatman smear tab | 100 | Methanol |
| O’Malley et al. (115) | DTBP | Cotton swab | — | Ethanol |
| OSHA (116) | Misc | Glass fiber filters | 100 | Wet/dry |
| Misc | Whatman smear tabs | 100 | Wet/dry |
| Rappe et al. (117) | PCBs | Kleenex | 200 | Dry |
| Stephens (118) | PCBs | Kleenex | 200 | Wet |
| Vostal et al. (163) | Lead | Commercial wet wipes | 900 | Ethanol |
| Liou et al. (89) | Multiple | Template wipe | 100 | Dry/wet |
| HUD (119) | Lead | Commercial wet | 100 | Wet |

Abbreviations: DTBP, 2,4-diter-butylphenol; HUD, Housing and Urban Development; MDA, 4,4’-methyleneedianiline; Misc, miscellaneous; NA, not available; OSHA, Occupational Safety and Health Administration; PCBs, polychlorinated biphenyls; TCDD, 2,3,7,8-tetrachlorodibenzo-p-dioxin; Adapted from McArthur (87).
material already on the hand, and the transfer of particles to the hand is a complex process. Loading was dependent on many variables, including dampness of the skin and surface roughness. However, they did show that the maximum uptake for coarse Arizona road dust, 20–40 µm in diameter, was greatest within the first 5–10 repeated presses. Our studies showed that the human hand collected the particles < 250 µm with the greatest efficiency. Calibration studies for the EL sampler and the Rodes et al. laboratory studies (124) indicated that a number of issues still need to be resolved for future versions of surface samplers that mimic pick-up by hands. Included is how to collect information necessary to describe the activities and frequency of activities that led to contact with toxicant laden surfaces.

The Department of Housing and Urban Development (HUD) has used the baby wipe sampler as the method of choice for comparisons of lead values to a household clearance standard after remediation or control of lead sources in a home (125). It is simpler to use than the LWW sampler and is designed specifically to determine the surface loading of lead and other heavy metals in micrograms per square centimeter or micrograms per square foot (119). The simplicity of the HUD device allows it to be used to compare the levels of lead detected with a clearance value for lead to define when a home is “clean.” It is based on the amount of lead picked up by the sampler in the defined space. This is accomplished by using a wetted material (Little Ones, Baby Wipes Lightly Scented; Kmart Corporation, Troy, MI) and moving it around a defined template area-sampling surface. However, there is no control for operator pressure, and the wetting agent can remove materials from deep below the surface being sampled.

Farfel et al. (102) compared the HUD wipe sampler with an HVS3 vacuum sampler, and Rich et al. (105) compared the HUD wipe sampler with the LWW sampler. However, each was a performance test and did not test what the collected sample represents. However, Rich et al. (105) used a calibrated LWW method for comparison with the HUD sampler.

The comparison completed by Rich et al. (105) using the LWW and the HUD sampler showed that the HUD sampler would consistently pick up more lead than the LWW sampler. Analysis of the data suggested that the HUD sampler scours the surface more deeply, but it also had much more variability and a much higher detection limit than the LWW. Paustenbach (26) noted in his evaluation of surface samplers that operator pressure is a major concern when attempting to assess the reproducibility and utility of surface samplers for characterization of exposure.

Vacuum samplers. The vacuum cleaner samplers pose their own series of problems for dust collection. First, not all vacuum cleaners are alike. In fact, they can have very different designs and particle collection characteristics and particle retention characteristics. However, vacuum cleaner bag samples have been used for many years to collect dust samples. One vacuum sampler designed specifically to collect house dust was the HVS3. It collects the dust from a floor using various cyclone collectors for particles > 5 µm and a final HEPA filter for collection of smaller particles. It has been used successfully in a number of programs to collect house dust from rugs and bare floors (86,103). There are also a number of rug samplers that use different types of mini-vacuum cleaners.

Farfel et al. (103) compared the efficiency of two vacuum systems, the HVS3 and the CAPS (Comprehensive Abatement Performance Study) cyclone. They used three types of dust: a large-diameter dust sample (25–2,000 µm) from the U.S. EPA, an intermediate-size Buffalo River sediment (39–149 µm; NIST-SRM-2704; National Institute of Standards and Technology, Gaithersburg, MD), and a small size (0.5–44 µm) sample of USP talc (United States Pharmacopia, Rockville, MD). This study was a first and important step toward standardization of soils for vacuum cleaner calibration comparisons.

The vacuum cleaner, as a tool for forensic analysis of house dust, has advantages over the wipe sampler. It can provide the investigator with a large quantity of mass, which can subsequently be used to detect the range of toxicants and toxicant levels in the dust. Further, the concentration or loading can be compared to the values used for residential cleanup of soil or for determining potential exposures that may cause a specific health
The simplest sampling method is to collect dust from residential vacuum cleaner bags (126,127). Although lacking the precision of systematic designed vacuum sampling methods, this is an effective tool for gross identification of contaminant levels in homes. Thus, it is an important screening tool for identifying high-end exposures. To improve the applicability of the results from standard vacuum bags, information needs to be gathered on the characteristics and activities in the homes. Included would be the age of house, age of carpet, when the vacuum cleaner bag was replaced, residential construction activities, ventilation patterns, and cleaning patterns.

Obviously, more precision can be obtained using a standardized vacuum cleaner that collects a documented size fraction of dust and that vacuums a standard section of carpet (17,89,128). Finally, the vacuum cleaner method of dust collection has also been used in conjunction with polyurethane foam (PUF) rollers to characterize the distribution of pesticides on surfaces and at various levels in a carpet (129) or with the LWW sampler to examine surfaces in a home (89,92).

Other samplers. The attic dust sampler, the surface deposition plate, mats, microscope slide plates, or sticky tapes and rollers are devices that can provide quantitative information on the particle size and composition of material that have been generated indoors or outdoors. The sampling device is deposited on a surface and left undisturbed for a specific amount of time. In the case of all but the attic dust sample, the location, timing, and the duration of the sample are selected by the investigator. Analytically this is helpful because some types of events and sources influencing the deposited material can be qualitatively or quantitatively identified using survey tools before or during sample collection. During the use of a deposition collector, the accuracy of source identification can be improved if videotape is running throughout the sampling period, but this is difficult to achieve because of logistics. A prospective sampling strategy can be tailored to the toxicant or the source of concern.

An attic sample can represent the long-term dynamic accumulation of material that has been influenced for many years by the natural movement of air, penetration of the dust indoors, and the eventual deposition of dust on many surfaces. Most samples are collected and placed in a bag using plastic scoops or a wisk broom (96,97). The major concern is that important household activities and home maintenance may disturb the attic dust (e.g., roof replacement, renovations).

Farfel et al. (94) placed floor mats in residential entryways to examine the movement of material into and out of a residence. Floor mats can be used to quantify the dust or toxicant levels that are tracked into the residence from outdoors and accumulate on the mat over a specified period of time. This study focused on comparing the collecting efficiency of two vacuum cleaners for removal of deposited dust, and they concluded that older homes appeared to yield higher lead loading and this was due to higher lead concentrations in the deposited dust. They also did comparative testing of the devices for standard reference material that contained lead. Included were two NIST samples and a bag house sample. Farfel et al. (94) indicated that the mat collection technique needs further evaluation if it is to be used to estimate dust deposition rates. They stated that it is essential to determine which type of mat should be used, and eventually what the collected data can represent to assess exposure and eventually risk for toxicants deposited on the mat left in an entry way. Finally, they made the important point that we need to find better methodologies for determining incremental dust deposition in the home.

Edwards et al. (55) used a deposition plate to determine the amount of particulate matter that settled on a deposition plate over the course of 30 days. The plates were located on a flat surface at 0.3 and 1.5 m off the ground. The deposited particles were measured by image analyzer to determine the particle size distributions by height and time of the year. The results revealed significant differences in deposition based on particle size alone, with large numbers of deposited particles in the winter and greater deposited mass in the summer. The latter was due to the influence of large particle deposition. The lower height samples had more particles in the summer-time, but the size of the deposited particles was smaller during the winter.

Pellizzari et al. (52) and Bonanno et al. (53) used entrance mats to examine the tracking of metals indoors during the NHEXAS. Farfel et al. (94) used mats to collect and analyze samples for lead. In each case, the sample was left in the doorway for a specific period of time, approximately 3–4 weeks, to collect an incremental amount of mass with or without contamination. Nishioka et al. (25) used SOF-WICK pads to collect pesticides deposited on lawns. They also used PUF rollers and cotton gauze (25,130).

Finally, surface soil samples, which would be taken from the top 1–3 cm of soil, are another source of information on potential residential exposures. The purpose is to obtain the scientific data for estimates of dermal contact and incidental ingestion by collecting soil in the yard that has a high potential for sustained human contact either currently or in the future. Thus, surface soil can be analyzed to determine if the soil is the source of toxicants of concern. The technique usually involves selecting a Teflon or other non-background laden pan and using an appropriate brush to sweep the surface soil into the pan. After a prescribed area is sampled, the material is placed in a sealed bag and sent out for analysis. This technique has been used by numerous investigators for a variety of toxicants. As an example, archived soil samples collected over 36 years in the United Kingdom at two experimental stations were recently analyzed for organochlorine pesticides. The results showed decreases in levels over time, and the results indicated the peak soil concentrations coincided with peak usage in the 1960s (131).

In addition to examining a rug or other surfaces for individual particles, elements, and other physical or chemical constituents, one can obtain quantitative and qualitative information on bacteria, molds, spores, and other viable particles. For the material present on individual surfaces, it is possible to transfer material to slides for analysis or to petri dishes filled with an agar for colony growth and identification. In addition, the investigator can place a sampler such as a petri dish to collect biologicals that can deposit on surfaces (132).

Other issues. The materials that can be subjected to analysis are far ranging in types and characteristics. From the standpoint of forensic or exposure research analyses this is exceptionally good because one can address a number of different types of exposure issues and source receptor relationships. The major concern is 2-fold: how to triage the analytic opportunities and how to collect enough of the appropriate sample type to detect the toxicant of concern. Collection issues affect our ability to measure heavy metals, organic compounds, asbestos, other inorganic species, radionuclides, and viable and nonviable biological materials. Particle size and morphology are also helpful for identifying and addressing many residential exposure issues. The work of Rudel et al. (80) piloted a methodology to examine hormonally active agents and animal mammary carcinogens in house dust. This advanced the use of
dust to measure phenolic compounds, phthalates, in addition to the more common toxicants PAHs and pesticides. The samples were collected using a “Mighty-Mite” vacuum cleaner. Unfortunately, the samples were not differentiated by surface, so the results can only be considered “whole” residence samples. One observation was that house dust provides a record of past exposure and can increase understanding of potential exposures due to the use of commercial products. It should be examined for pre- and postevent issues.

The work of Molhave et al. (78) focused on microorganisms and allergens, as well as aldehydes and the basic components of dust. Their goal was to demonstrate that dust is a source of airborne particles found in a workplace. However, an important aspect of the experiment was the use of the contents of a typical vacuum cleaner bag as the source of material for analyses. Thus, for forensic and screening experiments, it appears logical to examine the content of the basic vacuum cleaner bag as a first step in a triage for conducting exposure and assessments in a home for research and/or forensic purposes to select the major contaminants of concern.

Dust not only is present on surfaces and carpets, but it can transfer to people’s hands. This is of particular concern for children who exhibit mouthing behaviors. Methods for collecting dust from hands can be as simple as wiping the hands with a moist towelette or paper or rinsing the hand with water, weak acid, or solvent, or placing the hand on the surface of a culture medium. The type of method used depends on the aspect of the dust that is of interest. Several investigators have used hand wipes for collection of dust. Vostal et al. (83) used commercial hand wipes, Wash’n Dri Towelettes, to collect dust from the hands of children and from windowsill or wall surfaces. Duggan et al. (133) used another commercial hand wipe, Wet Ones, to collect hand wipe samples and also evaluated a range of commercially available wipe media. The hand wipe method was found to be simple and reproducible when the contaminant of concern was lead. The efficacy of the method in collection of lead has been demonstrated for both house dust and playground dust, with the first wipe collecting between 50 and 70% of the lead on hands. The use of the same method for collection of hand wipes and environmental samples makes it easy to compare the measures obtained from the two sources.

The challenge in using hand wipes to measure total dust on children’s hands is that the sampling medium is not easily preweighed and therefore a mass cannot be directly calculated. A way to get around this problem is to back-calculate mass using the measurements of metals or pesticide found in size fractions of dust that will strongly adhere to hands and other locations on the skin. This method was recently used with metals (90).

Hand rinses with water or rubbing alcohol have been used to collect dust when the materials of concern are pesticides (87,120,130,134,135). The disadvantage of hand rinses are their messiness in collection and the labor-intensive reduction methods necessary before analysis. It is possible, however, through careful evaporation of the rinsate, to obtain a measure of dust mass on the hand (133,135).

Assessment of Data Obtained from Residential Dust

Information on activities and lifestyles needed to assess residential exposure to dust and other media requires the application of survey instruments to record home and occupant histories before determining the exact location of dust sampling. For instance, an exposure-based investigation first requires information on the who, what, when, where, and why associated with a particular problem. This information can be gathered in a variety of ways, including questionnaires, videotaping, diaries, and interviews (88). Each of these tools can be modified to address problems in specific situations. For example, questionnaires exist that can be used to identify sources, characterize the physical structure of the house and the activities in the house, identify external and internal source types, health status of occupants, and product uses. Daily diaries and videotaping are survey instruments that can be used to get prospective information on the likelihood and frequency of potential contact with chemical, physical, and biological agents.

Fairly simple questionnaires can be used to evaluate the “dustiness” of homes. Rough estimates of the amount of visible dust on a 3-point scale have been found to associate well with blood lead levels in children (136) and with dust levels on windowsills (121). These scaled indicators of dust have also been associated with chromium loadings on windowsills and urine chromium of residents near chrome waste sites.

Questions posed to householders that address cleaning habits, ventilation and heating practices, residential construction, and presence of pets and children in the home can also provide valuable information related to the dust levels and composition found in homes.

Videotaping is a new and useful tool in exposure assessment. It provides an unobstructed and noncumbersome approach to examining passive or active contact with an environmental toxicant (137,138). With new video technology, miniaturized cameras can now be deployed that do not require the shadowing of an individual subject by a technician carrying a video camera pack. Such data can be an extremely valuable addition to forensic analyses of exposure because digitalized results from the videotaping can help focus attention on specific locations, sources, and times that require sampling. The videotape also helps reduce the amount of speculation about what sources or conditions or activities lead to contact. Videotaping is an evolving tool and is extremely valuable for characterizing exposure to materials and contaminants that have multimedia sources and multiple exposure pathways.

In the area of inhalation exposure, for example, certain volatile organic solvents can be emitted and accumulate in the ambient air, by drinking and showering water, and by specific products used in or around a home or workplace (2,139). Videotaping provides an objective measure of the types and locations of duration of potential or actual contact with volatile toxicants.

Each home or building environment has general dynamic conditions associated with the structure’s inherent characteristics and the lifestyle of the occupants. The variables that can influence the accumulation, movement, and removal of materials present in a rug include type of housing, the season, age and condition of housing, furnishings, insulation, behavior and activities of occupants, construction, and location with respect to outdoor sources. Gathering information on the large number of variables that can affect a home environment is a daunting task; however, many variables will have a long-term and consistent influence on the levels of materials present in the carpet. Thus, specific variables may provide information on levels for materials present, which can range in composition from starch through heavy metals. Other variables will reflect deposits caused by specific events, accidents, planned activities, discrete changes in lifestyle, and changes in the number of occupants.

Freeman et al. (89) have shown that food that is frequently dropped on the floor accumulates toxicants, and children who play and on contaminated surfaces will transfer toxicants to the foods they eat when they do not wash their hands before handling food. A recent analysis by Bonanno et al. (53) of the NHEXAS data obtained in a statistically representative population within six Midwestern states has demonstrated that the home characteristics and activities are good indicators of the potential for toxicant contamination in house dust and indoor air. This is an important observation that can be used in the design and implementation of field studies because such information can help with defining the triage for a dust analysis and sampling strategy.

Hunt et al. (77) and Adgate and colleagues (140,141) were among the first groups to attempt multiple source signature
identification and differentiation (142). In the case of Hunt et al. (77) and Adgate et al. (140), the analyses were done for lead present in residential carpets and other surfaces, respectively. Because lead can be emitted by many types of sources, just identifying the presence and levels of lead in a carpet will only provide information on the potential level of contamination. The lead levels alone, however, will not provide quantitative source signatures even if only one source is supported. Fortunately, individual sources or source types may yield different emission profiles that include other elements and materials beyond lead.

Adgate et al. (140) used the statistical technique of chemical mass balance and a suite of elements detected in the rug along with the measured lead to match with patterns or fingerprints of major source types that contributed to the lead levels in the rug. The results were then used to apportion the percentage of the lead mass that would be contributed by each source. This approach had been used successfully for detailed analysis of air pollutant sources and the origins of crushed rock using a series of home samples from Jersey City, New Jersey. Adgate et al. (140) identified interior house paint, street dust, and yard dust as the major ultimate or proximate sources of the lead. This information can be used to identify interim and final remediation strategies for children affected by high lead in blood (> 10 µg/dL). Hunt et al. (77) performed studies using the scanning electron microscope energy dispersive X-ray analysis technique to identify sources and characterize samples for lead and then applied cluster analysis on the dust from six households in London. The authors concluded that road dust was the highest contributor to lead, with paints and soils being the next largest contributors. For other multiple source toxicants, similar techniques can be used (e.g., isotope ratios), while in other cases just the measurement of an individual compound of concern or unique tracer that easily identifies a source would be adequate. For the latter, a good example would be pesticides and herbicides that are sprayed inside or outside of the home. Normally, each has one unique or distinctive active ingredient, which provides a distinct marker or tracer for the source (e.g., backyard and farm applications or residential applications). In the case of viable particles, environmental conditions of the home or a particular room, rug, or carpet may offer an opportunity for breeding more material (e.g., bacteria and mold). Thus, visible material may not just be an indication of moisture problems, but also the source of viable biological particles that can cause specific health outcomes.

Brown et al. (72) were able to discern the source of lead infiltrating a home by microscopic analysis. In this case, the morphology of the particles found in residences, neighborhood soils, and industries were compared to the morphology of the particles found in uncovered piles of electric furnace flue dust that had been dumped under and around the house. The analyses showed that the lead-laden particles had been transported from the piles into the surrounding area, which resulted in lead and cadmium contamination. The difference between the above techniques was that the studies of Hunt et al. (77) and Brown et al. (72) were qualitative, while the study of Adgate et al. (140) was quantitative. However, each was able to differentiate sources in residences and other structures.

House dust sampling has also been found to be useful in analyzing the success of interim or long-term mitigation or remediation strategies used for indoor or outdoor sources that affect a residence. Examples include the study of Roberts et al. (50) that showed the reduction in lead loading on rugs after repeated vacuuming. A similar study conducted by Farfel and Chisolm (143) in Baltimore tried to link the reductions in lead loading after cleaning with changes in blood lead. This approach was also used in Jersey City, New Jersey, by Rhoads et al. (21), who showed an average drop in blood lead in children of 17% after a year-long cleaning intervention. The reductions in lead loading over the course of the year for both the cases and controls were measured using surface wipe sampling and rug or carpet sampling. The approach used by Rhoads et al. (21) involved collecting initial and final samples with the LLW and a vacuum sampler that were used to define the reduction in potential exposure (89,122).

Application of Household Dust to Exposure/Source Characterization

Chromium exposure characterization. A series of studies was conducted to determine the influence of chromate production waste sites and waste site remediation of residential chromium concentrations and the exposure of individuals who lived near the waste sites (20,51,144–146). These studies were conducted in Jersey City over a 10-year period. During the first study (144), several environmental measures of exposure were used: soil samples close to residences, indoor and outdoor air samples, indoor vacuum samples, and indoor surface wipe samples. Vacuum samples were collected from 200-cm² floor sections by the front and back entryways onto cellulose fiber filters using a Gast diaphragm pump. Dust wipe samples were collected with a 50-cm² template using the LWW sampler.

Among the indoor measures in the first study (144), air levels of chromium were fairly low, typically between 0.5 and 1 ng/cm³. The air samples were correlated with the smoking patterns of the residents and were not indicative of exposure from outdoor sources. In contrast, wipe samples ranged from nondetectable to 320 ng/cm², with a median household level of 24 ng/cm². The wipe sample surface loadings were significantly higher in homes near chromium waste sites than in control homes. Vacuum dust chromium loadings were much lower than what was found in wipe samples (1.9 ng/cm²). Although the loadings obtained by the two house dust collection methods were different, the concentration of chromium in dust samples was similar.

Impact of human activities on chromium in house dust could be observed in this first study (144). The concentration of chromium in wipe dust samples (micrograms per gram) on windowills tended to be higher than levels detected on interior surfaces such as refrigerator tops and bookshelves. Concentrations in house dust were also associated with residential cleaning habits such as when dusting or sweeping was performed last. House cleaning was seen to have several effects: it could episodically increase airborne chromium after vacuuming and reduce surface chromium levels after dusting.

In follow-up studies performed in conjunction with the New Jersey Department of Health Chromium Medical Surveillance project, the LWW wipe sampler was used to collect house dust specifically from windowills (51,146). Again it was demonstrated that chromium dust loadings were associated with house-cleaning practices as reported by homeowners. Chromium loadings were lowest in homes that had been cleaned recently and greatest in homes that had been cleaned more than one week before to the dust sampling visit.

Dust loadings and chromium loadings in the home were affected by the absence of doormats, an important means of reducing dust tracked in from outdoors. In addition, chromium dust levels in homes were also associated with urine chromium of the residents.

Table 5. Chromium concentrations in house dust before and after remediation for median homes that initially had high, medium, and low concentrations using the LWW sampler.

| Category of home | Chromium (µg/g) | Initial visit | Final visit |
|------------------|----------------|--------------|-------------|
| Low Cr           | Mean ± SD      | 58 ± 33      | 54 ± 106    |
|                  | Median         | 48           | 51          |
|                  | n              | 7            | 5           |
| Medium Cr        | Mean ± SD      | 248 ± 76     | 66 ± 77     |
|                  | Median         | 245          | 34          |
|                  | n              | 8            | 6           |
| High Cr          | Mean ± SD      | 782 ± 331    | 55 ± 60     |
|                  | Median         | 239          | 50          |
|                  | n              | 8            | 7           |

Adapted from Freeman et al. (59).
suggesting either inhalation of resuspended or ingestion of the chromium in dust (51,146).

Evaluation of homes after remediation of the waste sites found the LWW wipe sampler to be an effective tool for evaluating reduction in chromium concentrations following the removal of the chromate waste site source (51,146). Concentrations of chromium showed significant reductions from levels preceding site remediation (Table 5).

**Forensic analysis and evaluation of exposure to biologies and other particles using microscopic techniques.** Environmental forensic microscopy can play a useful role in the investigation of indoor exposure concerns. Like criminal forensic analysis of trace evidence, microscopical analyses of environmental samples can provide information about the identity and possible source of contaminants. As described earlier, fungal spores, pollens, skin cells, synthetic fibers, natural fibers, and animal hairs are prominent in the dust of many buildings. Air pollution particles deposited in the building (e.g., metallic fragments), soots, and building material debris may also be present. Thus, as discussed in the previous section, identification of these particles by microscopical analysis is useful in determining whether the source of the possible irritants is in the building or comes from outside. Polarized light microscopy using dispersion staining and microchemical tests is used to identify a range of particles found in airborne and settled dust samples, from fungus to pollens to combustion products to fibers (synthetic and natural) (147). The scanning electron microscope with X-ray elemental analysis capability (SEM/EDS) augments the light microscopic examination with information about the elemental composition of materials. In the automated mode, the SEM/EDS can examine thousands of particles in a sample of dust. The information about size and elemental composition of the particles can be organized to give a comprehensive inventory of a dust sample that is useful in determining the source. The transmission electron microscope with electron diffraction and X-ray analysis capabilities is used to identify very small particles (including soots), < 1 μm in diameter. Infrared microscopy is used to identify particles made of organic molecules such as polymers and pharmaceuticals. The following case studies illustrate the use of environmental forensic microscopy in the investigation of indoor contaminants.

**Sample collection.** In the following, samples of airborne particulate were collected with polycarbonate filter cassettes attached to area pumps. Samples of surface dust were collected by dry wiping, adhesive lift, or vacuuming techniques depending on the amount of dust present and/or the question that needed to be answered (148). If the dust layer was substantial, a sample may have been collected with a clean spatula and placed in a clean plastic container. Lesser amounts were collected with a microvacuum constructed with an air sampling cassette and nozzle or by wiping with a plastic glove or with a plastic bag. Adhesive lifts were used in special circumstances where surfaces might have been damaged easily.

**Microscopy equipment.** Millette and Few (148) used several types of microscopy for particle identification:
- Stereomicroscopy using a Zeiss Stemi 2000 stereomicroscope having a magnification range from 6.5x to 47x
- Polarized light microscopy including microchemical tests using an Olympus BH-2 polarized light microscope with a magnification of 40–1,000x
- SEM using a JEOL 6400 coupled with an X-ray EDS Noran Voyager system
- Transmission electron microscopy (TEM) using a JEOL 1200, 100 kV scanning transmission electron microscope (STEM), equipped with a Noran EDS X-ray analysis system or a JEOL 2010, 200 kV TEM with an X-ray analysis system
- Infrared microscopy or, as it is more properly known, Fourier transform infrared microspectrophotometry (micro-FTIR) utilizing a Perkin-Elmer Auto Image System coupled to a Series 2000 FTIR.

All samples were first examined by stereomicroscopy and in most cases the particles were analyzed by polarized light microscopy. Depending on what was found, portions of the sample were analyzed by SEM, TEM, or infrared microscopy. Methods for the microscopy of particles can be found in the Particle Atlas (149).

**Case studies.** Applications of microscopic techniques as an approach to understanding issues of exposure intensity and source identification are provided in the following three case studies. One is an example of outdoor contamination causing indoor problems. The other two are related to microbiological issues related to moisture and dust mites, respectively.

**Case 1.** Black ghosting areas were found throughout a new house in Michigan. The black areas formed readily on plastic surfaces such as kitchenware, drapery rods, and medicine containers in medicine cabinets. Microscopic examination of the black material on a piece of drapery rod by TEM and X-ray elemental analysis showed that the black material was soot particles (Figure 2). The particles were consistent with carbon soot from paraffin burning, but candles were not used in the residence. Information gathered by surveying the residents showed that paraffin logs had been used during one period of time, but not concurrently with the black ghosting problem. Further investigation showed that there was a backdraft from the fireplace drawing air down the chimney into the residence. Once the chimney flue was closed, the problem went away. Apparently, the soot forming on the residence plastic materials was coming from the carbon soot deposited on the chimney.

**Case 2.** In a Georgia governmental office building, black and white particles were found several days in a row on a desk located under an air system duct grate in the late fall. Samples were collected by an adhesive lift. Light and scanning electron microscopy showed that the particles were black clusters of fungal spores and white fragments of galvanized metal. Apparently moisture accumulated in the ductwork during the summer season.

**Figure 2.** Soot particles as seen by TEM.

**Figure 3.** Mite found in dust as seen by polarized light microscopy.

**Figure 4.** Mite found in reference insulation dust as seen by SEM.
and caused a growth of fungal material as well as a corrosion of the galvanized metal in the air supply system. When the atmosphere in the building turned cooler and drier, the particles were released from the duct system.

**Case 3.** In a convent in Oklahoma, residents complained of eye irritation and general itchiness. Glass fibers from the duct insulation were suspected of causing the problem. Air samples collected on polycarbonate filters and examined by light microscopy did not show the presence of glass fibers or other particles that are considered irritants. Analysis of particles associated with the duct insulation sent as a reference showed a high concentration of mites (Figures 3 and 4). Additional testing for mite antigens was recommended.

Finally, the microscopical analysis of particles has been a fundamental activity of the law enforcement forensic community for many years. The French detective Locard published several classic articles in the 1930s about the microscopical study of dusts to determine their origin (67). As these case studies have shown, microscopy using light, electron, and infrared microscopes can provide helpful information in characterizing a suspected exposure to environmental contaminants.

**Conclusions**

Dust in the home has traditionally been considered as a nuisance and material that must be removed by a vacuum cleaner or broom. Thus, historically little time was spent sampling or analyzing the material for contaminants. Lead exposure changed that philosophy. This review indicates that we have come a long way in determining the uses of house dust and residential samples to identify sources of indoor contaminants and to provide improved estimates of residential total human exposure. Research and applications of dust analyses in recent years demonstrate that we can take meaningful samples for the detection of a variety of chemical, physical, and biological toxicants. The challenge for the future is to continue the evolution of reliable techniques for wipe, surface, and vacuum samples. These are necessary to improve qualitative determinations of surface loading and dust concentrations. Finally, the efforts to establish performance evaluations and determine what a dust sample represents must become part of the process for development and selection of samples for use in research, regulating, and forensic investigations.

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Review • Dust exposure assessment and source characterization

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