Association between short-term exposure to sulfur dioxide and carbon monoxide and ischemic heart disease and non-accidental death in Changsha city, China

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Abstract

Background
To investigate the effects of short-term exposure to sulfur dioxide (SO₂) and carbon monoxide (CO) in the central and southern China areas on ischemic heart disease (IHD) and non-accidental deaths.

Method
We investigated the associations between short-term exposure to SO₂ and CO in a city in south-central China and IHD and non-accidental death using a time-series design and generalized additive models with up to a 5-day lag adjusting for day of the week, temperature, air pressure, wind speed, and relative humidity. The relative risks of IHD and non-accidental death per 10-unit increase in SO₂ and CO were derived from zero to five days in single-pollutant models.

Results
Between 2016 and 2018, a total of 10,507 IHD and 44,070 non-accidental deaths were identified. The largest significant relative risk for IHD death was lag 02 for both SO₂ (1.080; 95% confidence interval: 1.075–1.084) and CO (5.297; 95% confidence interval: 5.177–5.418) in single-pollutant models. A significant association was shown at all lag multiple-day moving averages. Two-pollutant models identified an association between SO₂ and mortality when adjusting for CO. In stratified analyses, SO₂ exhibited a stronger association with death during the cold season, while CO exhibited a stronger association with mortality from IHD during the warm season. The risk of death was more robust in the elderly for both pollutants, but was greater in men for CO and in women for SO₂.
Conclusions
Overall, we found an association between short-term exposure to low-level SO₂ and CO and the risk of IHD and non-accidental death.

1. Introduction
Globally, estimated years of life lost from ischemic heart disease (IHD) increased by 20.9% (19.0–22.9) between 1990 and 2007, and by a further 17.3% (15.4–19.0) between 2007 and 2017 [1]. IHD has become the second leading cause of death in China [2]; various clinical, epidemiological, and toxicological studies have provided evidence that ambient air pollution likely contributes to the development and exacerbation of IHD [3–6]. Ambient air pollution causes the deaths of an estimated 4.2 million individuals each year [7], and IHD accounts for 25% of deaths related to air pollution [8]. Sulfur dioxide (SO₂) and carbon monoxide (CO) have both been associated with a higher risk of heart failure and death [9–11].

SO₂ exerts adverse health effects through systemic inflammation and oxidative stress [12, 13]; in systemic circulation, SO₂ can enhance the risk of cardiorespiratory mortality and morbidity [14], coronary heart disease [15], and IHD [14, 16]. However, most research in developing nations has focused on particulate matter and not on SO₂ pollution. China is the leading SO₂ emitter in the world, with the average for annual-mean SO₂ levels in 338 Chinese cites reported as 3–87 μg/m³ in 2015 [17], well beyond the ranges in developed countries [18]. However, with the efforts made by the Chinese government in recent years to improve the ecological environment, SO₂ pollution has been largely mitigated and the average annual-mean SO₂ levels in 338 Chinese cites reported as 13 and 11 μg/m³ in 2018 and 2019, respectively [19]. Is SO₂ at lower concentrations associated with the deaths in China? The problem is worthy of exploring.

Ambient CO could cause myocardial ischemia and rhythm disturbance at lower concentrations [20], increase the number of strokes and enhance the risk of IHD mortality at short-term exposure [11, 21]. As the largest developing country in the world, China is undergoing unprecedented advances in urban motorization. As a result, ambient CO pollution has become a serious environmental issue in many cities in China where the number of motor vehicles has increased rapidly over the past few decades. As we all known, the incomplete combustion of fossil fuels in motor vehicles was mainly producing CO. However, compared with the ample evidence supporting the coherence and plausibility of the association between CO and mortality in the USA and Europe [22, 23], few studies have explored these associations in China, where air pollution characteristics, meteorological conditions, and socioeconomic patterns are different from those developed countries.

Non-accidental deaths consistent with the law of life or disease development, is considerable to estimate the association between short-term exposure to air pollutants and daily deaths, and is often used as the compare group with the disease being studied [24]. Although epidemiological studies in developed nations have provided evidence of an association between ambient SO₂ and CO pollution and mortality, fewer studies have been conducted on this topic in China. Thus, there is a pressing need to evaluate the association between the current exposure level of SO₂/CO and mortality from IHD. The present study was a time-series analysis of the effects of short-term exposure to SO₂ and CO in the urban districts of Changsha city in southern-central China on IHD and non-accidental deaths during the period of 2016 to 2018, to provide evidence for exploring the mortality risks associated with low concentrations exposure of SO₂ and CO in China.
2. Data and methods

2.1. Study area

Changsha is the capital city of Hunan Province (28°12’N, 112°59’E), with an area of 11,819 km² and a population of 7.9 million. Five urban districts (Furong, Kaifu, Tianxin, Yuelu, and Yuhua), are included in Changsha, totaling an area of approximately 1,216 km² (Fig 1). The population was approximately 3.5 million in 2019.

2.2. Daily death counts, air pollution, and meteorological data

The personal information of deaths between January 1, 2016 and December 31, 2018 were obtained from the National Mortality Surveillance System, operated by the Center for Disease Control and Prevention of Hunan province. The information of each death was including birth and death dates, sex, age, cause of death and other demographic factors. Causes of death were assigned from death certificates and were coded by the International Classification of Diseases 10th Edition (ICD10). We extracted and analyzed the information of deaths during the study period and districts by searching I20–I25 for IHD and A00-R99 for non-accidental deaths with the ICD-10 codes.

The air pollution data during the study period were collected by 10 monitoring stations in the five urban districts of Changsha and provided by the Changsha Environmental Protection Bureau. The concentrations of daily 24-h average SO₂ (μg/m³) and CO (mg/m³) were used as the exposure. The concentrations of SO₂ and CO included in the model were the average of the measurements of the 10 monitoring stations. These stations were consistent with the site selection requirements of national air monitoring stations, and they were located away from industrial sources, major roads, or residential sources of emissions including coal, waste, or oil. Thus, our results reflect the background urban air pollution level in Changsha rather than local sources of pollution such as traffic or industrial combustion [25]. K-nearest neighbor imputation for the outliers and few missing values of the air pollution data was implemented, and it was replaced by the average of the 3 days before and after [26].

Meteorological data, including temperature, air pressure, wind speed, and relative humidity during the study period were obtained from Changsha Meteorological Bureau. There were no missing data.

The study protocol was approved by the Ethic Review Committee of Changsha Environment Protection College and carried out in accordance with the principles of the Declaration
of Helsinki. All data on patient death records was fully anonymized prior to the researchers accessing them.

2.3. Statistical analysis

Mean ± standard deviation and quartile were calculated for the descriptive analysis of variables. Spearman’s rank correlation was used to evaluate the association between meteorological factors and air pollution. Daily data for concentrations of SO$_2$ and CO and weather conditions were pooled together for the same time to match the daily IHD and non-accidental deaths. A longitudinal time-series design was conducted to evaluate the associations between the short-term exposure to SO$_2$/CO and daily IHD and non-accidental deaths. A quasi-Poisson regression (quasi-likelihood) in generalized additive models was used to estimate the association between short-term exposure to SO$_2$ and CO and daily IHD and non-accidental deaths. Smoothing spline functions were applied to control the effects of confounding factors such as secular trends, meteorological factors, and day of the week.

We used the model $\log[E(Y_t)] = \alpha + \text{DOM} + \beta X_t + s(\text{time}, \text{df}) + s(Z_t, \text{df})$, where $t$ is the observation date; $Y_t$ is the daily death count on day $t$, $E(Y_t)$ is the predicted death count on day $t$; $\alpha$ is the intercept; DOM is the dummy variable for day of the week, $\beta$ is a coefficient in the regression model, $X_t$ is the concentration of pollutants on day $t$; $s$ is the natural cubic spline function; $Z_t$ is the meteorological data on day $t$ and df is degree of freedom.

In the model, $s(\text{time}, \text{df})$ and $s(Z_t, \text{df})$ denote the smoothing spline function for nonlinear variables which represent the calendar time and meteorological factor (daily average temperature, air pressure, wind speed, relative humidity) on day $t$, respectively. The df was selected according to the minimum Akaike Information Criterion. We applied single-air pollutant models to examine the effects of SO$_2$/CO on IHD and non-accidental deaths. Meanwhile, we assessed the lag effects of both single-day (distributed lag: lag0-lag5) and multiple-day moving averages (moving average lag: lag01-lag05). In the single-day lag models, lag0-lag5 mean the concentrations of air pollutants of the current day (lag0) and the previous several days (from lag1 to lag5). In multi-day lag models, lag05, for example, means a 6-day moving average pollutant. The two-air pollutant model was used to test the effect of one pollutant on IHD and non-accidental deaths with the other pollutant adjusted in the model. The lag effects of both single-day and multiple-day moving averages were also assessed.

The results are expressed as the relative risk (RR) of IHD and non-accidental deaths for every 10 ug/m$^3$ or 10 mg/m$^3$ increase in the concentrations of SO$_2$ and CO, respectively. The exposure relationships between SO$_2$/CO concentrations in different seasons and the log-relative risk of mortality were examined graphically by replacing the linear term of the SO$_2$/CO concentration with a smoothing function. The predicted values of the smoothing term, i.e. log-RR, were then plotted against the SO$_2$/CO concentrations (Fig 2).

Subgroup analyses were conducted for age (<65 and ≥65 years), sex (male and female), season (May to October [warm] and November to April [cold]) [9, 26, 27] SPSS 19.0 software (IBM SPSS, Armonk, NY, USA) was used for data input, and R statistical software (version 3.2.3, R Foundation for Statistical Computing, Vienna, Australia) was used for data analysis. P<0.05 was considered as statistically significant.

3. Results

3.1 Descriptive statistics

Table 1 shows the descriptive statistics of daily deaths, meteorological data, and air pollutant levels between January 1, 2016 and December 31, 2018. During the study period (1097 days), the total number of the IHD and non-accidental deaths was 10507 and 44070, respectively (9
IHD and 40 non-accidental deaths per day). The IHD daily death counts were higher in males and the elderly (≥65 years) compared with the females and the people less than 65 years. Average daily concentrations of SO\(_2\) ranged from 3.80 to 57.00 μg/m\(^3\) (mean: 12.96 μg/m\(^3\)). Average daily concentrations of CO ranged from 0.42 to 2.10 mg/m\(^3\) (mean: 0.85 mg/m\(^3\)). The concentration of the two pollutants was higher during the cold season \((P < 0.05)\). The average air pressure, temperature, relative humidity, and wind speed were 1001.33 ± 8.87 h Pa, 17.64 ± 8.73 °C, 79.27 ± 14.22%, and 2.63 ± 1.43 m/s, respectively.

3.2. Spearman correlation

S1 Table shows the daily mean values of the Spearman correlation coefficients between air pollutant levels and meteorological conditions in Changsha during the study period. SO\(_2\) and CO levels were moderately correlated with each other \((0.384, P < 0.01)\). Wind speed was weakly negatively correlated with SO\(_2\) and CO levels \((-0.253 and -0.210 respectively, \(P < 0.01)\). Air pressure was moderately correlated with CO levels \((0.367, P < 0.01)\).

3.3. Generalized additive modeling

Table 2 lists RR values and 95% confidence intervals (CIs) of IHD and non-accidental deaths associated with 10-unit increase in pollutant concentrations in lags ranging from zero to 5 days in single-pollutant models. For IHD deaths, the significant RRs observed for SO\(_2\) were for lag 1–2 and lag 01–05, while for CO, the significant RRs were for lag 0–3 and lag 01–03. The largest significant RR was Lag02 for both SO\(_2\) (RR: 1.080; 95%CI: 1.075–1.084) and CO (RR: 5.297; 95%CI: 5.177–5.418). The associations of SO\(_2\) and CO with mortality were more pronounced on moving-average days. Table 3 lists the association of air pollutant levels with IHD and non-accidental death stratified by sex, age group, and season. For IHD deaths, CO levels were associated with death more among men \((RR: 5.517; 95%CI: 5.381–5.654)\) and the elderly \((RR: 3.551; 95%CI: 3.439–3.664)\), and the association was stronger during the warm season \((RR: 7.319; 95%CI: 7.708–7.560)\). However, SO\(_2\) was statistically significant correlated with the warm season \((RR: 0.905; 95%CI: 0.895–0.914)\).

For non-accidental deaths, the significant RRs observed for SO\(_2\) were for lag 1–2 and lag 01–05, while for CO, the significant RRs were for lag 0, 1, 5 and lag 01–02. The largest
significant RRs in single-pollutant models were lag 02 for SO\textsubscript{2} [1.059 (95%CI: 1.057–1.062)] and lag 01 for CO [4.012 (95% CI: 3.900–4.124)] in Table 2. SO\textsubscript{2} levels were associated with non-accidental deaths among the elderly (RR: 1.042; 95%CI: 1.039–1.045); meanwhile, CO levels were associated with death more among men (RR: 4.495; 95%CI: 4.431–4.559), the elderly (RR: 3.072; 95%CI: 3.011–3.132) and the cold season (RR: 3.066; 95%CI: 3.003–3.129) in Table 3.

Fig 2 presents the exposure-response relationships between air pollutant levels and same-day deaths from IHD in the single-pollutant models during the warm season, the cold season, and annually. The exposure-response curves associated with SO\textsubscript{2} both during the cold season

### Table 1. Daily deaths, air pollutant levels, and meteorological data in Changsha, between 2016 and 2018.

|                          | Mean     | SD       | Min     | 25th    | 50th    | 75th    | Max       | Total    |
|--------------------------|----------|----------|---------|---------|---------|---------|-----------|----------|
| Daily death counts (N)   |          |          |         |         |         |         |           |          |
| Non-accidental deaths    | 40.21    | 10.7     | 14      | 33      | 39      | 45.25   | 119       | 44070    |
| IHD deaths               | 9.59     | 3.97     | 1       | 7       | 9       | 12      | 38        | 10507    |
| Death counts by gender (N) |        |          |         |         |         |         |           |          |
| Male                     | 23.45    | 6.88     | 7       | 19      | 23      | 27      | 67        | 5654     |
| Female                   | 16.76    | 5.63     | 4       | 13      | 16      | 20      | 52        | 4853     |
| Death counts by age (N)  |          |          |         |         |         |         |           |          |
| 0–65 years               | 9.20     | 3.28     | 0       | 7       | 9       | 11      | 20        | 1584     |
| ≥65 years                | 31.01    | 9.35     | 11      | 25      | 30      | 35      | 104       | 8923     |
| SO\textsubscript{2} (μg/m\textsuperscript{3}) |          |          |         |         |         |         |           |          |
| All year                 | 12.96    | 6.66     | 3.80    | 8.20    | 11.80   | 16.20   | 57.00     | -        |
| Warm season              | 11.64    | 4.86     | 4.00    | 8.00    | 10.60   | 14.40   | 32.40     | -        |
| Cold season              | 14.30    | 7.86     | 3.80    | 8.80    | 13.00   | 17.60   | 57.00     | -        |
| CO (mg/m\textsuperscript{3}) |          |          |         |         |         |         |           |          |
| All year                 | 0.85     | 0.23     | 0.42    | 0.68    | 0.80    | 0.98    | 2.10      | -        |
| Warm season              | 0.74     | 0.14     | 0.46    | 0.64    | 0.72    | 0.82    | 1.24      | -        |
| Cold season              | 0.96     | 0.25     | 0.42    | 0.80    | 0.94    | 1.10    | 2.10      | -        |
| Air pressure (h Pa)      |          |          |         |         |         |         |           |          |
| All year                 | 1001.33  | 8.87     | 983.10  | 993.80  | 1000.8  | 1008.00 | 1029.72   | -        |
| Warm season              | 995.21   | 6.01     | 983.10  | 990.55  | 994.20  | 999.13  | 1015.12   | -        |
| Cold season              | 1007.54  | 6.71     | 990.50  | 1003.07 | 1007.50 | 1012.59 | 1029.72   | -        |
| Temperature (°C)         |          |          |         |         |         |         |           |          |
| All year                 | 17.64    | 8.73     | -2.80   | 10.14   | 18.30   | 25.00   | 32.70     | -        |
| Warm season              | 24.53    | 4.83     | 9.75    | 21.40   | 24.92   | 28.41   | 32.70     | -        |
| Cold season              | 10.66    | 5.74     | -2.80   | 6.40    | 10.12   | 14.90   | 25.30     | -        |
| Relative humidity (%)    |          |          |         |         |         |         |           |          |
| All year                 | 79.27    | 14.22    | 35.50   | 69.5    | 80.75   | 91.00   | 100.00    | -        |
| Warm season              | 80.24    | 12.01    | 49.50   | 71.19   | 81.00   | 89.76   | 100.00    | -        |
| Cold season              | 78.28    | 16.08    | 35.50   | 68.19   | 80.75   | 92.76   | 100.00    | -        |
| Wind speed (m/s)         |          |          |         |         |         |         |           |          |
| All year                 | 2.63     | 1.43     | 0.00    | 1.58    | 2.43    | 3.54    | 8.45      | -        |
| Warm season              | 2.52     | 1.38     | 0.00    | 1.50    | 2.30    | 3.41    | 8.45      | -        |
| Cold season              | 2.73     | 1.48     | 0.00    | 1.60    | 2.52    | 3.66    | 7.38      | -        |

IHD, ischemic heart disease; CO, carbon monoxide; Max, maximum; SD, standard deviation; SO\textsubscript{2}, sulfur dioxide; Warm season, May to October; Cold season, November-April.
and annually and with CO during the warm season exhibited similar positive linear relationships. IHD deaths increased notably with incremental additions of CO levels, especially when exposure concentrations were low. Other descending-shaped curves tended to become nonlinear at higher CO concentrations during the cold season and annually. Fig 3 shows the association between pollutant concentrations and daily death counts in the two-pollutant models. For IHD deaths, SO$_2$ increased risk significantly from lag 0 to lag 05 but decreased it from lag 1 to lag 4, when CO was included in the model. CO did not increase IHD deaths risk only at lag 3 and lag 5 when adjusting for SO$_2$. For non-accidental deaths, SO$_2$ and CO decreased its risk significantly from lag 1–5 and lag 02–05, when adjusting for the other air pollutant in the model.

### Table 2. Relative risk and 95% confidence intervals of daily deaths associated with a 10-unit increase in the levels of sulfur dioxide (SO$_2$) and carbon monoxide (CO) for various lag days in single-pollutant models.

| Lag   | SO$_2$                  | CO                  |
|-------|-------------------------|---------------------|
|       | Ischemic heart disease deaths |
| Lag0  | 1.012(1.008–1.016)      | 3.256(3.152–3.359)$^*$ |
| Lag1  | 1.059(1.055–1.062)$^{**}$ | 3.023(2.924–3.121)$^*$ |
| Lag2  | 1.051(1.047–1.054)$^{**}$ | 3.451(3.352–3.550)$^*$ |
| Lag3  | 1.016(1.013–1.020)      | 0.845(0.743–0.946)   |
| Lag4  | 1.020(1.017–1.023)      | 1.615(1.512–1.718)   |
| Lag5  | 1.014(1.011–1.017)      | 0.680(0.576–0.783)   |
| Lag01 | 1.058(1.054–1.063)$^*$  | 4.012(3.900–4.124)$^*$ |
| Lag02 | 1.080(1.075–1.084)$^{**}$ | 5.297(5.177–5.418)$^{**}$ |
| Lag03 | 1.074(1.069–1.079)$^{**}$ | 3.779(3.650–3.908)$^*$ |
| Lag04 | 1.074(1.068–1.079)$^{**}$ | 3.821(3.683–3.958)   |
| Lag05 | 1.074(1.068–1.080)$^*$  | 2.925(2.780–3.071)   |
|       | Non-accidental deaths   |
| Lag0  | 1.026(1.024–1.028)$^*$  | 2.977(2.923–3.030)$^{**}$ |
| Lag1  | 1.039(1.037–1.041)$^{**}$ | 1.980(1.929–2.031)$^{**}$ |
| Lag2  | 1.032(1.030–1.033)$^{**}$ | 1.271(1.221–1.321)   |
| Lag3  | 1.009(1.008–1.011)      | 0.853(0.803–0.903)   |
| Lag4  | 0.749(0.698–0.800)      | 0.993(0.992–0.993)   |
| Lag5  | 0.990(0.989–0.992)      | 0.527(0.476–0.578)$^*$ |
| Lag01 | 1.058(1.054–1.063)$^{**}$ | 4.012(3.900–4.124)$^{**}$ |
| Lag02 | 1.059(1.057–1.062)$^{**}$ | 2.494(2.432–2.557)$^{**}$ |
| Lag03 | 1.053(1.051–1.056)$^{**}$ | 1.913(1.847–1.980)   |
| Lag04 | 1.043(1.040–1.046)$^{**}$ | 1.568(1.498–1.638)   |
| Lag05 | 1.034(1.031–1.037)$^*$  | 1.209(1.135–1.283)   |

$^*$p < 0.05;
$^{**}$p < 0.01.

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4. Discussion

In this population-based time-series study, we identified 10,507 IHD and 44,070 non-accidental deaths in Changsha between 2016 and 2018. CO levels were associated with mortality to a greater extent than SO$_2$, especially for IHD. Our findings suggest that the association of SO$_2$ with death from IHD is greater during the cold season and that of CO is greater during the warm season. In addition, the risk was greater in the elderly for both pollutants and in men for CO. Conversely, the risk of overall non-accidental death attributable to SO$_2$ levels was more robust in women. To the best of our knowledge, this is the first comprehensive study of the acute effect of both SO$_2$ and CO on mortality from IHD in China.
The main sources of SO$_2$ and CO in Changsha come from vehicle exhaust, dining fumes, coal-fired heating, and industrial emissions. SO$_2$ and CO levels in Changsha are higher during the cold season, when use of coal for energy is more frequent. Mean levels of SO$_2$ and CO in our study were higher than those reported in Wuhan, China and in developed nations [5, 28, 29], but lower than those reported in Shenzhen [30], Beijing [31], and Iran [32]. Nowadays, the China government has been working hard to tackle air pollution by coming up with policies and measures when founding the economic development has been achieved at the expense of the environment. Therefore, the concentration of all air pollutions is starting to decrease [33].

We found lagged associations between SO$_2$ and CO with IHD and non-accidental deaths. For each 10-mg/m$^3$ increase in CO, the RR peaked at a lag of 0–2 days. A nationwide time-series analysis in 272 major cities in China also found that associations between short-term exposure to CO and cardiovascular disease mortality were strongest during the first two lag days [21]. Our results are also similar to a previous estimate in a project conducted in 19 European cities reporting that the 2-day mean CO concentration was associated with an increase in cardiovascular disease mortality [22]. We found that the maximal impact of SO$_2$ occurred at a lag of 0–5 days for IHD deaths and 0–2 days for non-accidental deaths. A study in Hong Kong reported a positive correlation between SO$_2$ levels and IHD death after a 1-day lag [34], while a study in Wuhan estimated that SO$_2$ was associated with non-accidental deaths at lag0 and lag1 [27]. The cumulative lag effects of SO$_2$ and CO exhibited increasing trends for both IHD and non-accidental deaths. Investigating the associations between air pollution and disease mortality after delays may clarify their relationships and suggest the disease prevention advice [32].

The curves for SO$_2$ to IHD deaths in this study indicated that ambient SO$_2$ increased IHD deaths slowly during the cold season but inversely during the warm season, which was confirmed in the stratified analyses. These findings are concordant with published reports of seasonal variations in mortality [35]. SO$_2$ effect estimates were more pronounced during the cold season in five urban districts of Changzhou, China [9], but another study reported that SO$_2$ was linked to deaths from atherosclerotic heart disease during the warm seasons in Canada [36]. This shows that effects vary by study design, medical condition, pollutant level, and the presence of other pollutants. Temporal variations of vasoconstriction and higher cholesterol and triglyceride concentrations during the cold period could change mortality rates [37]. What is more, the presence of pathogens that cause respiratory infections may also be relevant to the seasonal variation of mortality [38].
We found that ambient CO levels were associated with slow increases in IHD deaths during the warm season; during the cold season, the association between CO levels and IHD mortality first increased and then decreased. Overall, low levels of CO appeared to contribute to IHD deaths while high levels appeared to decrease at high IHD deaths. A study conducted in 126 US urban counties between 1999 and 2005 also found evidence of an association between short-term exposure to ambient CO and risk of IHD hospitalizations well below current US regulatory standards [23]. At low levels, CO causes tissue hypoxia resulting from the binding of CO to hemoglobin in the blood [39]. Reduced oxygen-carrying capacity of hemoglobin predisposes toward cardiac ischemia in persons with coronary artery disease, and exacerbation of cardiovascular symptoms in persons with coronary heart or lung disease. However, we did not
found the association between high levels of CO and adverse health effects reported in other studies [40]. A systematic review and meta-analysis based on 2,748 articles published between 1948 and 2014 observed the strongest association between CO concentration and mortality in studies from low- and middle-income countries [41]. Pathways other than carboxyhemoglobin formation may cause the toxicity of CO, which is produced by incomplete combustion of hydrocarbons. In the urban areas, the largest source is vehicle exhaust emission, followed by industrial process, central heating and fires. CO concentrations are temporally heterogeneous, with lower concentrations and a greater influence on IHD deaths observed during the warm season, indicating that lower-level CO pollution may actually increase the public health burden. Residents of cities with warmer climate may engage in more outdoor activities, promoting individual exposure to CO. However, one study reported that a 10-mg/m³ increase in CO would increase IHD mortality by 21.1% after a 1-day lag [11].

We found a significant effect of SO₂ and CO on IHD and non-accidental deaths for all ages, but the association was stronger for the elderly (≥65 years), a known high-risk population [27, 42]. Higher risk in the elderly may be attributable to cumulative toxic effects from long-term exposure to ambient pollution and to concomitant cardiovascular morbidity [36]. We found that men were more at risk for IHD and non-accidental deaths associated with CO exposure than women, perhaps because men are more likely to work outdoors [43]. Conversely, women were more at risk for non-accidental deaths associated with SO₂ exposure, consistent with reports that SO₂ was only associated with non-accidental deaths in women [27, 44]. We could not identify the mechanism or underlying factors responsible for the sex disparity using a time-series analysis. Therefore, a more accurate exposure assessment or toxicological studies design should be leveraged to explain this phenomenon.

The study showed the association of SO₂ with deaths did not weaken after adjusting for CO levels. A multicenter European study conducted in seven countries reported that the association of SO₂ with IHD in individuals aged < 65 years was not modified by CO [14]. However, a study conducted in five urban districts of Changzhou City, China, showed an association between SO₂ levels and ischemic stroke mortality when adjusting for CO in two-pollutant models [9]. In the study of Wang et al., they reported that the association of SO₂ exposure with mortality was robust when adjusting for CO [13]. This suggests that SO₂ may play an independent role in triggering ischemic cardiac events. In contrast, the association between CO levels and mortality was modified by SO₂, indicating the collinearity of the two pollutants but also the persistence of the SO₂ effect on mortality. In China, a study conducted in four cities located in the Pearl River Delta observed strong confounding effects of particulate matter and nitrogen dioxide on cardiovascular mortality, in association with exposure to ambient CO [39]. Consistent with our findings, some studies reported that CO was associated with a decreased or non-significant risk of death after adjusting for the other pollutants [45, 46]. However, a study conducted in the 26 largest cities in China showed that the associations of CO with daily hospital admissions for all-cause and cardiovascular diseases were significantly robust to adjustment for co-pollutants [40], perhaps because of a combined effect of gaseous pollutants on cardio-cerebrovascular disease. Exposure to ambient levels of CO, SO₂, and other pollutants may strengthen or weaken their individual effects. Therefore, toxicological and population studies of exposure to air pollution mixtures may identify the biological mechanisms through which pollutants exert adverse health effects.

However, the limitations in our exposure assessment protocol should be noted in interpreting the results. Firstly, we relied on routine measurements from 10 fixed-site monitoring stations instead of more accurate measurement based on the individuals’ residence. Thus, we were not able to quantify the exposure of individuals precisely. This may lead to errors in the exposure measurement and underestimation of the effects. Second, we did not consider CO/
SO₂ exposure indoor sources. For us, no indoor air monitoring data is available, and future studies may consider that studying the effect of indoor air pollution on IHD deaths or the proportion of indoor air pollution on mortality in the overall air pollution effect. Thirdly, due to the ecological study design which is the use of aggregated data, concerns regarding residual confounding and ecological fallacy remain. Therefore, the results of this study cannot be generalized to the individual level and need to be interpreted as providing an etiological hypothesis. Moreover, the data regarding IHD cases and air pollutant levels were collected from only one city, and it is difficult to extrapolate the results to other areas in China. Further national or multicity studies are required.

The present study also has strengths. First, this study is one of few to report an association between SO₂ and CO and IHD in a provincial capital city exposure to slight to moderate levels of pollution. Second, our study consisted of a large sample of 10,507 and 44,070 IHD and non-accidental deaths, respectively, and there were no missing data meteorological and air pollution data for the study period, strongly supporting the statistical findings. Finally, the data were acquired from reliable sources and the weather monitoring stations covered the entire urban area of Changsha, representing the pollution situation well. The findings of this study can provide evidence to explore the health risks associated with low concentrations exposure of SO₂ and CO in south-central China.

5. Conclusions

In conclusion, we observed significant associations between short-term exposure to relatively low SO₂ and CO and increased daily risk of IHD and non-accidental death in the urban districts of Changsha city. Short-term exposure in the cold season to SO₂ and warm season to CO showed increased concentration-dependent associations with daily IHD deaths. The men have greater risk of death for CO exposure, while the women for SO₂, and the elderly for both pollutants. Our study provides evidence to explore the health risks associated with low concentrations exposure of SO₂ and CO in China.

Supporting information

S1 Table. Spearman’s rank correlation between air pollutants and meteorological factors in Changsha, China (2016–2018).

(DOCX)

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