In situ biosurfactant production and hydrocarbon removal by *Pseudomonas putida* CB-100 in bioaugmented and biostimulated oil-contaminated soil

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Abstract

In situ biosurfactant (rhamnolipid) production by *Pseudomonas putida* CB-100 was achieved during a bioaugmented and biostimulated treatment to remove hydrocarbons from aged contaminated soil from oil well drilling operations. Rhamnolipid production and contaminant removal were determined for several treatments of irradiated and non-irradiated soils: nutrient addition (nitrogen and phosphorus), *P. putida* addition, and addition of both (*P. putida* and nutrients). The results were compared against a control treatment that consisted of adding only sterilized water to the soils. In treatment with native microorganisms (non-irradiated soils) supplemented with *P. putida*, the removal of total petroleum hydrocarbons (TPH) was 40.6%, the rhamnolipid production was 1.54 mg/kg, and a surface tension of 64 mN/m was observed as well as a negative correlation (R = -0.54; p < 0.019) between TPH concentration (mg/kg) and surface tension (mN/m). When both bacteria and nutrients were involved, TPH levels were lowered to 33.7%, and biosurfactant production and surface tension were 2.03 mg/kg and 67.3 mN/m, respectively. In irradiated soil treated with *P. putida*, TPH removal was 24.5% with rhamnolipid generation of 1.79 mg/kg and 65.6 mN/m of surface tension, and a correlation between bacterial growth and biosurfactant production (R = -0.64; p < 0.009) was observed. When the nutrients and *P. putida* were added, TPH removal was 61.1%, 1.85 mg/kg of biosurfactants were produced, and the surface tension was 55.6 mN/m. In summary, in irradiated and non-irradiated soils, in situ rhamnolipid production by *P. putida* enhanced TPH decontamination of the soil.

Key words: bioremediation, irradiated soil, total petroleum hydrocarbons, rhamnolipids, *P. putida*.

Introduction

At present, Mexico has severe water, soil, and air pollution problems due to contamination by both organic and inorganic compounds. Total petroleum hydrocarbons (TPH) are considered to be high-priority compounds that should be removed from polluted areas in Mexico (SEMARNAT, 2008). TPH are a mix of pollutants, and in soil, they represent a risk to the health of people and ecosystems (DOF, 2005). Some of these pollutants have low solubility in water and are, therefore, difficult to remove from polluted environments.

For such pollutants, the use of synthetic surface agents has been suggested. These agents enhance the pollutants solubility’s (Volkering et al., 1995), leading to their desorption from the soil, thereby making them more accessible to microbial degradation. Certain microorganisms are able to produce extracellular substances (biosurfactants) with similar characteristics to those of synthetic surfactants. Because they have the advantage of being highly biodegradable, these microbial compounds are attractive agents to use in the bioremediation field (Maier and Sobrén-Chávez, 2000).

*Pseudomonas putida* can produce rhamnolipids (Amézcua-Vega et al., 2004) and are able to remove phe-
nanthrene in liquid culture medium under certain culture conditions (Martínez-Toledo et al., 2006). The aim of this work was to measure biosurfactant production by \textit{P. putida} in a biostimulated and bioaugmented soil sample and to assess its ability to remove TPH from polluted soils.

Materials and Methods

Soil sampling and processing: The soil used in this study came from the Petrochemical Complex located in Poza Rica, Veracruz, Mexico, 20°30’6” N and 97°28’33” W. The complex is located in the hydrological region RH-27 “Tuxpan Nautla,” which includes the river basins of Nautla, Tecolutla, Cazones, and Tuxpan, as well as the Tamihua lagoons. According to the FAO/UNESCO/IRIS, sedimentary rocks are the predominant geological formation of the soil and Vertisols-Gleyic is the dominant type of soil. The area has a warm and wet climate with an annual mean temperature of 24.4 °C, annual precipitation of 1,472.4 mm, and a relative humidity between 76 and 80% (INEGI, 1997).

Waste from petroleum drilling operations, maintenance, and duct repair has been accumulating in this soil. For this study, the soil was sampled (Figure 1) at a depth of 50 cm at thirteen points (S.1 - S.13) and down to 2 m at six points (P.M.1 - P.M.6). The bulk samples were homogenized manually and sieved (2-cm mesh), then kept at 4 °C in polypropylene containers holding 200 kg. From these containers, 200 kg was taken and air-dried at room temperature, homogenized manually again, and 500 g was taken as a representative sample as reported elsewhere (Cline, 1994). Each sample was crushed with a mortar and pestle, passed through a standard sieve of 2-mm mesh size, and kept at 4 °C in glass amber bottles until its characterization and analysis. Gamma irradiation of the soil was performed at the Gamma Irradiation Department of the National Institute of Nuclear Research located in the state of Mexico. The soil was irradiated at 25 kGy in 2-kg polyethylene bags in 5-kg containers (ININ, 2010).

Physical and chemical characteristics of the soil: Humidity was measured by the gravimetric method (Gardner and Klute, 1982). The method used for total nitrogen (N) determination (Kjeldahl method) was that reported by Muñoz-Iniestra et al. (2000). Available phosphorus (P) was determined according to Bray and Kurtz in 1945 with modifications (Roldán-Martín et al., 2006). Organic matter was determined through the method of oxidation with chromic and sulfuric acids developed by Walkley and Black in 1947 (Muñoz-Iniestra et al., 2000). The pH was determined in a 1:10 (w/v) suspension of the soil in distilled water and measured with a Jenway Mod. 3020 potentiometer using epoxy electrodes (Muñoz-Iniestra et al., 2000).

After soil characterization, nutrients were balanced based on the biostimulation experiment. The nutrient sources were NH$_4$Cl (20 mg/g) and NaNO$_3$ (320 mg/g) for N and KH$_2$PO$_4$ (3.62 mg/g) and K$_2$HPO$_4$ (46.66 mg/g) for P, and these were added to obtain a carbon:nitrogen:phosphorus (C:N:P) ratio of 100:10:1 in the soil (Cookson, 1995). Water content was adjusted with sterilized water to 35%.

Inoculum preparation: Preparation for the bioaugmented treatment consisted of inoculation of \textit{P. putida} CB-100 into a culture medium (Martínez-Toledo and Rodríguez-Vázquez, 2010) containing (g/L): glucose (13.6), yeast extract (0.01), Fe$_2$SO$_4$ 7H$_2$O (0.0005), KH$_2$PO$_4$ (0.2), K$_2$HPO$_4$ (2), MgSO$_4$ 7H$_2$O (1), NaNO$_3$ (5), FeCl$_3$ 6H$_2$O (0.0049), KCl (1.1), NaCl (1.1), CaCl$_2$ 2H$_2$O (0.05), ZnSO$_4$ 7H$_2$O (0.006), MnSO$_4$ H$_2$O (0.006), CoCl$_2$ 6H$_2$O (0.0006), and CuSO$_4$ 5H$_2$O (0.0006). The pH was adjusted to 7 and incubation continued for 5 days at 150 rpm and 37 °C. After this time, colony-forming units (cfu) were measured to ensure that the number of \textit{P. putida} colonies was suitable for inoculation (at 10$^9$ cfu/mL) of the experimental samples. Analyses of N, P, and glucose levels in the inoculum were also performed.

Microcosm units: The microcosm units consisted of sterile serological bottles containing 20 g of soil (irradiated or non-irradiated) with three replicates of each one of the treatments that are mentioned in the followed paragraph. They were incubated at 37 °C. Destructive analyses were performed after 0, 2, 4, 6, 8, and 10 d of incubation. Aeration of the microcosm units was performed every 72 h under sterile conditions by shaking with a vortex for 1 min and leaving the lid open under sterile conditions for about 5 min. After 10 d of incubation, the microcosm units were established: 1) soil + sterilized water = control treatment; 2) soil + nutrients = biostimulated treatment (BiostT); 3) soil + \textit{P. putida} = bioaugmented treatment (BioauT); 4) soil + \textit{P. putida} + nutrients = bioaugmented-biostimulated treatment (BioauT-BiostT).

Colony counts: Heterotrophic microbial growth was monitored (bacteria and fungi). Soil microorganisms were
enumerate by plate counting methods (Clark, 1982) replicated three times. For bacterial quantification, tryptic soy agar (TSA) was used as a medium and incubation took place at 30 °C for 48 h. For fungal quantification, Bengal Rose agar plus 0.01% chloramphenicol was used and the plates were incubated for 7 d at 28 °C in a General Signal incubator. The microorganisms were reported as cfu per gram of dry soil.

Extraction of TPH: Extraction was achieved following the modified shaking extraction method of Schwab et al. (1999). Further, 1 g of the dry soil with 2 g of Na2SO4 and 5 mL of CH2Cl2 were placed in a conical centrifuge tube, vortexed for 1 min, and then centrifuged at 6,500 rpm for 20 min, and the supernatant was retained. This procedure was repeated three times starting from the addition of 5 mL of CH2Cl2. The organic extracts were combined and concentrated to dryness in a rotary evaporator (740 mbar and 40 °C); the residue was dissolved in C2Cl4 and filtered through a glass microfiber filter (Whatman GF/F). The efficiency of this shaking extraction method was previously evaluated and compared with Soxhlet extraction (Arce-Ortega et al., 2004; Banks et al., 2000).

Analysis of TPH by infrared spectrophotometry: The modified EPA 8440 (http://www.epa.gov/epaoswer/hazwaste/test/main.htm) was used for TPH quantification. A standard curve was generated using the EPA standard reference oil containing a mixture of isooctane, chlorobenzene, and hexadecane with ethylene tetrachloride added to reference oil containing a mixture of isooctane, chlorobenzene, and hexadecane with ethylene tetrachloride added to 100 mL. The curve was constructed with five points from 0 to 800 mg/L. Readings were made with an infrared spectrometer (Nicolet Mod. Nexus 470); adsorption was measured at 2,930 cm⁻¹ using a 10-mm Spectra-Tech Inc. IR Quartz TPH cell; data acquisition was performed with the EZ Omnic ESP 5.1 program.

Biosurfactant extraction from soils: The biosurfactants were extracted with an electrolyte solution composed of 10 mM KNO3, 10 mM Tris-HCl, and 3 mM NaN3 (used to prevent microbial contamination) at pH 7.0; a 1:2 (w/v) soil:solution relationship was used by a modified method (Noordman et al., 1994). Ten 10-mL solution samples were placed in 22 × 175 mm Pyrex tubes with Teflon lids and 5 g of soil (from the microcosm units). The tubes were left for 3 d under recirculating shaking at 240 rpm and ambient temperature (25-27 °C). Afterwards, they were left to rest for 2 d and then filtered through a 0.22-μm cellulose filter (Millipore). The liquid was used to measure surface tension and to assess the presence of rhamnolipids.

Surface tension measurements: All glass and polypropylene items used for this and for the rhamnolipid determination were acid-washed (0.1 N HNO3), rinsed with distilled water, and dried in a stove. Measurements of each biosurfactant extract (10 mL) were performed in triplicate. Surface tension was used as an indirect measure of biosurfactant production (Fox and Bala, 2000). The method is based on the maximum vertical force that is required to separate a stirrup or a ring from the surface of the liquid being examined and to draw up the film that has formed. A Kruss Processor Tensiometer K12, equipped with a ring, and a Kruss PolyScience digital temperature controller were used at 25 °C. After each measurement, the ring was cleaned with distilled water and then heated briefly above an ethyl acetate flame.

Biosurfactant assays: One milliliter of the biosurfactant extracted from the soil was used for the rhamnolipid analysis to assess the presence of rhamnose, the 6-deoxihexose method was used as reported by Rhaman et al. (2000a). The extract was mixed with 4.5 mL of concentrated H2SO4, vortexed for 2 min (these procedures were done in an ice bath in order to control the strongly exothermic reaction), and then warmed to 100 °C for 10 min (in a warm water bath). Afterwards, it was cooled to room temperature, 0.1 mL of freshly prepared of 3% thioglycolic acid (TGA) was added, and the mixture was shaken again and incubated for 3 h in darkness at ambient temperature. The absorbance was read immediately thereafter at 400 and 430 nm, and a standard curve of rhamnose that had received the same treatment as the samples was used for comparison. The formula employed to assess rhamnolipid content was that reported elsewhere (Rhaman et al., 2000a).

CO2 evolution: Heterotrophic activity was evaluated by an indirect method using CO2 evolution. Gas samples (2 mL) from the air spaces of the microcosms were collected with a 5-mL gas-tight syringe. A Gow-Mac 550 gas chromatograph equipped with a thermal conductivity detector and an Alltech CTRI stainless steel column was run under isothermal conditions at 25 °C; the injector was set at 40 °C and the detector at 100 °C. Helium was used as the carrier gas at a flow rate of 45 mL/min (Amézcua-Allieri et al., 2003). Data were processed with an integrative Gow-Mac. CO2 evolution was reported in milligrams of CO2 per gram of dry soil (mg CO2/g dm). The experimental period was determined by the instantaneous CO2 production in both irradiated soils and non-irradiated soils.

Statistical analyses

Analyses of variance (ANOVA), general linear model procedures (GLM), Pearsons correlations, and Tukeys multiple comparison test (all at a 0.05 significance level) were performed using SAS (Statistical Analysis Treatment) for Windows version 9.0.

Results

Polluted soil characterization

The results of the soil characterization are presented in Table 1. In petroleum-polluted soil, the main obstacle for bioremediation is the excess carbon content from hydrocarbon contamination, which affects the C:N:P ratio. Nevertheless, the soil pH and organic matter content were considered sufficient to support the bioremediation process (Cookson, 1995; Riser-Roberts, 1998).
Gamma irradiation of the soil

Gamma irradiation of the soil was performed and, by comparison with non-irradiated soil, the changes in the initial flora present and their ability to remove toxins and produce biosurfactants during the experimental period were determined. To investigate whether the initial TPH concentrations were altered by gamma irradiation, extractions were performed on irradiated soil samples (n = 3) and on non-irradiated soil samples (n = 3). The statistical analysis (ANOVA) revealed no significant difference (p < 0.61) between the TPH concentrations of the non-irradiated and irradiated soil samples; therefore, the TPH were not affected by gamma irradiation. However, gamma irradiation only destroys and inactivates part of the microbial population of the natural contaminated soil.

CO₂ evolution

The experiments were stopped based on the instantaneous CO₂ evolution observed in irradiated soil and non-irradiated soil treatments (Figure 2). In the irradiated soil: a total of 1.61 mg of cumulative CO₂/g dm were produced in 10 d with BioauT and, with BioauT-BiostT, 1.72 mg/g dm were produced; these two treatments produced similar values as did the control and BiostT. In the non-irradiated soil: with BioauT-BiostT, 2.95 mg of cumulative CO₂/g dm were produced in 10 d and, with BioauT, 1.87 mg/g dm were produced; the control and BiostT presented analogous results. A negative correlation was observed between P content and cumulative CO₂ concentration (R = -0.78, p < 0.0001). However, with BiostT the accumulation of CO₂ was less than with control treatments.

Microorganism activity

Bacterial growth (BacG)

In the irradiated soil, a quick reactivation of bacteria was observed with slight fluctuations during the experimental period (Figure 3A). In all experimental treatments, maximum cell growth was observed at day 6 (at values of approximately 2.6 x 10⁸ cfu/g dm); this microbial population could be resistant to gamma irradiation so they were reactivated under appropriate culture conditions. In the non-irradiated soil, the maximum cell growth in all treatments was observed at day 6 (Figure 3B). In the control treatment, bacteria were reactivated at day 4, and the maximum growth was 2.8 x 10⁸ cfu/g dm. With biostimulation, BacG fluctuated over time with a maximum of 2.1 x 10⁸ cfu/g dm. In BioauT, the maximum BacG was 5.6 x 10⁸ cfu/g dm, after which it declined. Finally, in BioauT-BiostT, bacterial numbers remained constant (maximum of 1.4 x 10⁸ cfu/g dm).

Fungal growth (FungG)

In the irradiated soil, the reactivation of fungi after gamma irradiation was lower than that of bacteria (Figure 4A). Fungal colony numbers (just the microorganisms that could be resistant to gamma irradiation and were reactivated under appropriate culture conditions) also fluctuated over time. At day 10, a maximum of 4.4 x 10⁵ cfu/g dm were observed under control conditions. With BiostT, faster reactivation was detected than with the control; 1.17 x 10⁵ cfu/g dm were observed at day 10. BioauT presented fluctuations in FungG with a maximum of 1.2 x 10⁵ cfu/g dm at day 6. In BioauT-BiostT, growth was slow, but on day 8 rapid growth was observed that continued until day

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Table 1 - Soil characterization.

| Parameter                        | Value       | Method                |
|----------------------------------|-------------|-----------------------|
| Moisture                         | 21.15%      | Gravimetric           |
| Water-holding capacity           | 14.4%       | Gravimetric           |
| pH                               | 7.14        | Potentiometric        |
| Total nitrogen                   | 0.1677%     | Kjeldahl              |
| Available phosphorus             | 1.823 mg/kg | Bray-Kurtz            |
| Organic matter                   | 6.30%       | Walkley-Black         |
| Total petroleum hydrocarbons     | 32.421 mg/kg|                      |
| Bacteria                         | 4.08 x 10⁶ cfu/g dm | Agar method |
| Fungi                            | 2.40 x 10⁵ cfu/g dm | Agar method |
| Textural composition             | Sand 31%, silt 34%, clay 35% | Bouyoucos |

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Figure 2 - Instantaneous CO₂ production in A) irradiated soil and B) non-irradiated soil.
10 (1.95 x 10^3 cfu/g dm). In non-irradiated soil, in the control, the colony numbers fluctuated over time, with the maximum count (9.2 x 10^5 cfu/g dm) observed at day 8 (Figure 4B). With BiostT, fungi grew during the first half of the experiment to 5.6 x 10^5 cfu/g dm. With BioauT, better growth was observed with a maximum of 1.0 x 10^6 cfu/g dm on day 4. In BioauT-BiostT, colony numbers fluctuated at the beginning of the experiment, but the maximum was observed on day 4 (5.3 x 10^5 cfu/g dm). Humidity measurements remained constant throughout the experimental period in all the experimental units, and this condition supported microbial activity in the soil. The pH ranged from 7 to 7.5 in all experimental treatments with irradiated soil; these conditions favored BacG. In non-irradiated soil the pH ranged from 6.9 to 8.0; these values sustained the interaction between the bacteria and fungi in the soil to enhance removal of contaminants.

Pearsons correlation analysis revealed that, in BioauT with either non-irradiated or irradiated soil, cumulative CO₂ evolution was significantly positively correlated with FungG (R = 0.64, p < 0.003 and R = 0.81, p < 0.007, respectively). In the irradiated soil, cumulative CO₂ evolution was correlated with BacG (R = 0.68, p < 0.003) in BioauT, and in BioauT-BiostT growth of bacteria and fungi was correlated with CO₂ evolution (R = 0.55, p < 0.048, R = 0.78, p < 0.002, respectively). In the non-irradiated soil, FungG was correlated with cumulative CO₂ (R = 0.649, p < 0.003) in BioauT, and in BioauT-BiostT, BacG was correlated with cumulative CO₂ evolution (R = -0.55, p < 0.016). Therefore, it could be inferred that CO₂ production in the experimental treatments could be attributed to the microbial activity in the soil (both irradiated and non-irradiated), but in the non-irradiated soil FungG played an important role into the microbial population of soil.

Nutrient consumption

Phosphorus (P)

Was the most highly consumed nutrient (Table 2). In the irradiated soil: P concentration ([P]) was depleted in both BiostT (80%) and BioauT-BiostT (76%) at day 10. In the non-irradiated soil: with BiostT, P was 74% consumed at day 10, and in BioauT-BiostT, P was 53% consumed at day 10.

Nitrogen (N)

In the control treatment and in BiosurfT with either irradiated or non-irradiated soil, N concentration ([N]) did not vary significantly during the experimental period (Table 2). In the irradiated soil: with BioauT, 35% of the N was consumed at day 2, and in BioauT-BiostT, N was consumed at the same rate (34%) on the same day. Pearsons correlation analysis indicated that [N] was correlated with BacG and rhamnolipid production in BioauT (R = -0.82, p < 0.0001 and R = 0.69, p < 0.003, respectively) and, in combined BioauT-BiostT, [N] was correlated with the decrease in surface tension (R = 0.55, p < 0.037). These results show that N was an important nutrient for BacG and biosurfactant production and hydrocarbon removal by P. putida 599.
factant production. In the non-irradiated soil: with BioauT-BiostT, 30% of the [N] was depleted at day 6. In this treatment, the statistical analyses indicated that [N] and [P] were positively correlated (R = 0.66, p < 0.002), and in BioauT, [N] was positively correlated with FungG (R = 0.64, p < 0.003) and negatively correlated with [P] (R = -0.64, p < 0.003). In this case N was the preferred nutrient of the fungal population.

**Table 3** - TPH concentrations (mg/kg) and removal (%) in the experimental treatments.

**TPH removal**

TPH removal was due to the presence of microorganisms (Table 3), and each treatment displayed a different behavior. Pollutant removal was quantified as the difference between the TPH values at the beginning and end of each experiment. ANOVA revealed that the experimental treatments had a significant effect (p < 0.0001) on TPH concentration (mg/kg). Multiple comparison tests revealed four

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**Table 2 - Nutrient concentrations in the experimental treatments.**

| Time (days) | Control IS | Biostimulated IS | Bioaugmented IS | BioauT-BiostT IS |
|-------------|------------|------------------|-----------------|-----------------|
| 0           | 1.44 ± 0.23| 30.67 ± 0.42     | 2.06 ± 0.42     | 27.12 ± 4.5     |
| 2           | 1.16 ± 0.31| 29.99 ± 6.02     | 1.32 ± 0.78     | 15.06 ± 0.0     |
| 4           | 1.23 ± 0.45| 21.20 ± 5.45     | 0.96 ± 0.15     | 22.35 ± 3.4     |
| 6           | 1.03 ± 0.09| 11.76 ± 0.36     | 2.50 ± 0.91     | 15.66 ± 1.5     |
| 8           | 1.64 ± 0.26| 11.71 ± 1.27     | 1.24 ± 0.41     | 8.81 ± 1.5      |
| 10          | 1.24 ± 0.38| 6.05 ± 1.25      | 1.06 ± 0.03     | 6.38 ± 3.25     |

**Nitrogen concentration (%)**

| Time (days) | IS NIS |
|-------------|--------|
| 0           | 0.16 ± 0.01 |
| 2           | 0.10 ± 0.00 |
| 4           | 0.11 ± 0.01 |
| 6           | 0.17 ± 0.01 |
| 8           | 0.16 ± 0.01 |
| 10          | 0.16 ± 0.01 |

I S = Irradiated soil; NIS = Non-irradiated soil; ± = Standard deviation.

**Table 3 - TPH concentrations (mg/kg) and removal (%) in the experimental treatments.**

| Time (days) | Control IS | Biostimulated IS | Bioaugmented IS | BioauT-BiostT IS |
|-------------|------------|------------------|-----------------|-----------------|
| 0           | 33,398 ± 8,444 (0.0%) | 30,204 ± 3,355 (0.0%) | 31,75 ± 5,017 (0.0%) | 35,966 ± 2,748 (0.0%) |
| 2           | 38,459 ± 2,574 (-15.1%) | 19,507 ± 5,311 (35.0%) | 13,833 ± 3,313 (56.4%) | 33,989 ± 1,230 (5.9%) |
| 4           | 35,768 ± 3,557 (-7.0%) | 18,853 ± 5,043 (37.5%) | 19,832 ± 4,153 (37.5%) | 15,607 ± 3,961 (56.6%) |
| 6           | 39,103 ± 3,906 (-17.0%) | 12,691 ± 2,135 (57.9%) | 22,300 ± 5,162 (29.7%) | 14,695 ± 3,980 (59.1%) |
| 8           | 37,187 ± 1,0037 (-11.3%) | 22,471 ± 6,081 (35.6%) | 20,679 ± 4,630 (34.9%) | 8,765 ± 1,539 (75.6%) |
| 10          | 31,470 ± 3,757 (5.8%) | 21,419 ± 5,600 (29.0%) | 24,020 ± 3,557 (24.5%) | 13,986 ± 1,832 (61.1%) |

Note: The negative values represent pollutants desorption and the positive values represent pollutant removal. IS = Irradiated soil; NIS = Non-irradiated soil; ± = Standard deviation.
groups; only the control treatment was significantly different from the other experimental treatments. In the irradiated soil, pollutant desorption was observed in the control treatment, but the best results were obtained with BioauT-BiostT: the initial TPH concentration was 35,966 mg/kg and the final TPH concentration was 13,986 mg/kg (61.1% removal). The remaining experimental treatments (BioauT and BiostT) produced fluctuations between removal and desorption of pollutants. Negative correlation was observed between TPH removal (%) and rhamnolipid production (mg/kg) \((R = -0.55, p < 0.033)\) in BiostT, but also a correlation between cumulative CO\(_2\) evolution and TPH concentration (mg/kg) was observed in BioauT-BiostT.

In the experimental treatments with non-irradiated soil, ANOVA revealed that the experimental treatments had a significant effect \((p < 0.0035)\) on TPH concentration (mg/kg). Tukey’s multiple correlation test \((\alpha = 0.05)\) revealed three different groups (control and biostimulated treatments were significantly different), but no significant difference was observed between bioaugmented and BioauT-BiostT treatments. Fluctuations were observed between desorption and pollutant removal. The control treatment exhibited similar behavior to BiostT: both showed similar fluctuations, but this treatment had the lowest removal efficiency (13.5%) whereas BioauT-BiostT produced 33.7% TPH removal. BioauT had the best pollutant removal efficiency of 40.6%; in addition, Pearson’s correlation test indicated that the surface tension decrease (mN/m) was negatively correlated with TPH concentration (mg/kg) \((R = -0.54, p < 0.0194)\). Figure 5 shows only the results of TPH removal (%) and rhamnolipid production (mg/kg) obtained for BioauT-BiostT with irradiated soil (Figure 5A) and non-irradiated soil (Figure 5B). In this figure it is possible to discern that, when the quantity of rhamnolipid increases, TPH removal also increases. Although the rhamnolipid production was similar, the TPH removal (%) was different in the two experimental treatments; perhaps this was due to the higher microbial population in non-irradiated soil and the possibility that some of these microorganisms use as an accessible carbon source the rhamnolipids produced by other microorganisms.

Surface tension

To monitor the presence of biosurfactants in the experimental treatments, the surface tension was assessed in extracts of all analyzed soil samples. An extraction solution with a surface tension of 75.2 mN/m was used, and this was considered to be the initial surface tension in all experimental units. In the experimental treatments with irradiated soil (Table 4), the greatest surface tension decrease (taken as the difference between the initial and the final surface tensions) was observed in BioauT-BiostT (19.6 mN/m) on day 10, and a similar result was observed in BiostT (18 mN/m) on the same day. In the non-irradiated soil, the greatest surface tension decrease (11.7 mN/m) was observed in BiostT at day 10, and the second best result was 10 mN/m in BioauT. Consequently, microbial populations with and without irradiation could produce surface active substances that promote changes in the surface tension as \(P.\ putida\) does.

Rhamnolipid content

Rhamnolipids were present in all treatments with irradiated and non-irradiated soil (Table 4). In the irradiated soil, again, the best treatment was BioauT-BiostT; it had the highest rhamnolipid content with a maximum of 2.9 mg/kg at day 8. This was followed by the control and bioaugmented treatments with 2.05 mg/kg at day 4; under the latter conditions, rhamnolipid production was positively correlated with [N] \((R = 0.69, p < 0.0031)\) and negatively correlated with BacG \((R = -0.64, p < 0.0094)\). The lowest production of rhamnolipids was obtained with BiostT (1.46 mg/kg) at day 6 and, under these conditions, a correlation was observed between BacG and rhamnolipid concentration \((R = -0.62, p < 0.0168)\).

In the non-irradiated soil, rhamnolipid production was 1.79 mg/kg in the control on the second day and, with biostimulation, it was 1.33 mg/kg on the last experimental day, which was the lowest level of rhamnolipid production. The results with BiostT were similar to the results obtained with the control treatment. BioauT-BiostT yielded the best result with 2.42 mg/kg rhamnolipids at day 4. In general, lower rhamnolipid production was observed with non-irra-
Table 4 - Indicators of biosurfactant production in the experimental treatments.

| Time (days) | Control | Bioaugmented | Bioau-Biost | Biostimulated |
|-------------|---------|--------------|-------------|---------------|
|             | IS      | NIS          | IS          | NIS           | IS            | NIS          | IS            | NIS           |
| 0           | 1.83 ± 0.07 | 1.79 ± 0.09  | 1.88 ± 0.06 | 1.6 ± 0.18    | 2.25 ± 0.21   | 1.70 ± 0.03  | 1.97 ± 0.24   | 1.70 ± 0.35  |
| 2           | 1.98 ± 0.19 | 1.79 ± 0.09  | 1.73 ± 0.18 | 1.97 ± 0.19   | 1.82 ± 0.14   | 1.69 ± 0.03  | 1.73 ± 0.20   | 2.04 ± 0.13  |
| 4           | 2.05 ± 0.05 | 1.73 ± 0.08  | 1.70 ± 0.04 | 1.50 ± 0.32   | 2.05 ± 0.12   | 1.69 ± 0.03  | 1.78 ± 0.25   | 2.42 ± 0.18  |
| 6           | 1.66 ± 0.05 | 1.74 ± 0.07  | 1.46 ± 0.28 | 1.71 ± 0.20   | 1.65 ± 0.00   | 1.61 ± 0.03  | 1.72 ± 0.22   | 1.86 ± 0.44  |
| 8           | 1.75 ± 0.14 | 1.59 ± 0.01  | 1.54 ± 0.16 | 1.41 ± 0.07   | 1.90 ± 0.02   | 1.54 ± 0.05  | 2.90 ± 0.03   | 2.21 ± 0.16  |
| 10          | 1.70 ± 0.14 | 1.61 ± 0.06  | 1.59 ± 0.19 | 1.33 ± 0.17   | 1.79 ± 0.15   | 1.54 ± 0.05  | 1.85 ± 0.08   | 2.03 ± 0.19  |

Initial surface tension = 75.2 mN/m; IS = Irradiated soil; NIS = Non-irradiated soil; ± = Standard deviation.

diated soil than with irradiated soil. In the biostimulated and Bioau-Biost treatments the rhamnolipid concentration was correlated with BacG (R = 0.50, p < 0.0548 and R = 0.67, p < 0.0043, respectively). The rhamnolipid production observed in this work could be attributed to BacG.

Discussion

The type of soil used in this work is characterized by aged pollutants with relatively high organic matter levels and a clay-like texture that could limit the biodegradation of the petroleum pollutants therein (Gardner and Klute, 1982). In these experiments, the pH ranges favored bacterial growth and ensured the establishment of P. putida in the soil (pH adjustments were not necessary). Humidity remained constant and indicated that the temperature employed was adequate.

Gamma irradiation applied to the soil destroyed some and inactivated temporarily others of the native microorganisms; it could be inferred that the latter kinds of microbes produce endospores that make them very resistant to irradiation. Therefore, the irradiated soil had very different strains as well as quantities of microorganisms compared with the non-irradiated soil (naturally contaminated soil).

In spite of the nutrient consumption in the biostimulated soil, the cumulative CO₂ evolution in this experiment was lower than that in the control treatment. Regardless, the nutrient ratio used in the current study (100:10:1, C:N:P) was sufficient to activate the autochthonous soil microorganisms and induce them to produce rhamnolipids. Our results are similar to those of other authors (Haddox, 1997; Liebeg and Cutright, 1999; Ruberto et al., 2003) who did not observe significant microbial growth in their BiostT, even though N and P were added to the soil at concentrations found in healthy soils. In this work, P was more rapidly depleted than N in the biostimulated and Bioau-Biost experimental treatments, but N was more important for biosurfactant production than P. A possible explanation for this could be that P was initially less available to the microorganisms in the soil (Silvia et al., 1999); thus, when it was added in solution to the soil, the microorganisms demonstrated a preference for this nutrient over the others. Total N did not show relevant changes in concentration in any of the experimental treatments. However, our results showed that, in the bioaugmented treatments with non-irradiated soil, N stimulated fungal growth and, in irradiated soil, this nutrient stimulated BacG and also favored rhamnolipid production.

The relative abundance of P. putida in the bioaugmented and Bioau-Biost experimental treatments and the cumulative CO₂ evolution revealed that P. putida was not inhibited by either the autochthonous microorganisms or the nutrients added (the added C:N:P ratio was adequate for P. putida growth). A possible explanation for this result is that the inoculum still contained low concentrations of macro- (N and P) and micronutrients, as well as glucose, which allowed the establishment of the microorganisms in the soil. Moreover, P. putida has advantages over other bacteria because it does not require an adaptation time (before or after soil inoculation) as do Pseudomonas spp. and Acinetobacter strain B-2-2 (Haddox, 1997; Ruberto et al., 2003). The fungal colony counts (maximum cell growth at day 6) and the cumulative CO₂ evolution in the BioauT with irradiated and non-irradiated soil, as well as the corre-
loration observed between fungal and bacterial growth, suggest that fungal growth was favored by the presence of *P. putida* (maximum cell growth at day 6), and that they could work synergistically; this finding agrees with findings in other studies (Bouchez et al., 1995; Aitken et al., 1998; Chávez-Gómez et al., 2003).

As the highest TPH removal was obtained in the experimental treatments with *P. putida*, these results confirm that the autochthonous microorganisms were incapable of pollutant removal (as can be seen in the results for the non-irradiated soils, control and BiosT), or alternatively that the nutrients did not increase the potential of the autochthonous degrading microorganisms, and this coincides with other reports (Nyman, 1999; Ruberto et al., 2003). These findings can be attributed to the different responses of each microorganism to the types and amounts of available nutrients. As in other reports (Trzesicka-Mlynarz and Ward, 1996; Madsen and Kristensen, 1997), in our experiments improvements in pollutant removal were only obtained in the presence of external inocula.

The surface tension changes in the soil extracts confirmed the presence of biosurfactants. The fluctuations in the surface tension changes and in the rhamnolipid concentrations could be due to a larger number of microorganisms that were reactivated in the non-irradiated soil consuming biosurfactants as co-substrates (Volkering et al., 1998). Moreover, the pollutant desorption observed during these experiments could be attributed to the biosurfactants (including rhamnolipids) produced by the microorganisms or to the disruption of cellular remnants in the soil, as in the report of Schippers et al. (2005). In other reports (Mackay and Gschwend, 2001; Hwang and Cutright, 2002), a similar phenomenon has been observed and was attributed mainly to cellular disruption in dissolved organic matter or in acid-soluble humic acids. Besides, similar surface tension decreases have been reported (Johnsen and Karlson, 2004), but in liquid culture with different microorganisms using polycyclic aromatic hydrocarbons as the carbon source. It has been suggested (Jordan et al., 1999) that biosurfactant production in soils is due to the need to make the nutrients available, and a characteristic of these soils is their low nutrient availability (mainly N, C, and P) due to the presence of pollutants. It can be inferred that biosurfactant production by microorganisms under these conditions is due to the same process of making the nutrients available and uses certain TPH as carbon sources. In the irradiated soil, the surface tension was smaller than in the non-irradiated soil. Nevertheless, it was possible to observe a correlation between the TPH concentration and surface tension in the bioaugmented non-irradiated soil.

The presence of microorganisms highly resistant to gamma irradiation can be inferred, as well as that of rhamnolipid producers; hence, it is likely that *P. putida* and/or *Pseudomonas* spp. were present in the polluted soil as autochthonous microorganisms. The main achievement of our work is that we have shown *P. putida* can adapt to this type of soil and remove pollutants through its rhamnolipid production; the rates of TPH removal were similar to those reported for other microorganisms (Ross et al., 1995; Rahman et al., 2002b) but in our work they were enhanced within a short time. This study demonstrated that *P. putida* (a non-pathogenic microorganism) is able to improve the removal of TPH from soil through its *in situ* rhamnolipid production.

**Conclusion**

The highest TPH removal was obtained in BioauT-BiosT with irradiated soil (modified microbial population), this treatment had the highest rhamnolipid production (2.90 mg/kg, at day 8); these low lipid concentrations together with nutrients and inoculated bacteria were effective in enhancing pollutant removal, which reached 75.6% at day 8. Besides it can be observed that N had a greater effect on biosurfactant production in the soil than P. Moreover, with non-irradiated soil (with native microorganisms), the rhamnolipid production was 2.42 mg/kg at day 4 but TPH removal was only 15.9%. Despite the similar amounts of rhamnolipid produced, *P. putida* was not able to remove TPH at the same rate; this could be attributed to the different types of strains that could grow with *P. putida*. Also *P. putida* (BiaugT) could enhance the TPH removal in the non-irradiated soil. The methodology employed in this work allowed us to determine the behavior of *P. putida* in aged polluted soil and their effectiveness in biosurfactant production and TPH removal. The results show this could be an attractive approach in field-scale bioremediation.

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