Aircraft events correspond with vocal behavior in a passerine

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Airports can affect birds by hindering acoustic communication. Here, we investigated the impacts of aircraft events on vocal behavior in wood thrush (Hylocichla mustelina) breeding one mile from an airport in Ithaca, NY, USA. We identified the number of wood thrush songs between 0500 and 0800 h at various distances from the airport and on days with various morning flight schedules. We also analyzed the number of sites from which birds sang during the peak of aircraft events (proxy of number of wood thrush). We found that birds sang more from 0600 to 0640 h when there were aircraft events during this period. This increased vocal behavior is likely explained by increased song output per individual wood thrush, rather than more wood thrush vocalizing. Increased song rate may negatively affect wood thrush fitness through increased energetic demands and/or time tradeoffs with other important behaviors, such as foraging. Identifying the noise thresholds associated with fitness costs (if any) and how different behavioral strategies (i.e. changing the pattern of vocalizations) may allow individuals to evade these costs would be useful for establishing conservation policy in breeding habitats used by passerines, such as the wood thrush.

Transportation networks continue to expand to accommodate the growing human population, causing anthropogenic noise to affect land area across the urban–rural gradient. Indeed, anthropogenic noise, such as noise from aircraft takeoffs, landings and flyovers, has even been found to double noise levels in the majority of protected areas in the United States. Noise profiles produced by aircraft are complex and broadband, due to many components simultaneously producing sound at both high and low frequencies (e.g., engine noise and airframe noise). Additionally, aircraft noise profiles manifest in a frequency range that is audible to many wildlife species (approx. 500–5000 Hz) and are relatively high intensity compared to other anthropogenic noise sources. Areas approximately 300 m from a 2-lane highway with relatively high traffic volume are exposed to approximately 47 dBA, whereas areas approximately 300 m from a jet take-off are exposed to approximately 100 dBA—a difference that is perceived by humans to be 32 times louder. The response of wildlife to aircraft noise is an area of current research, and a recent meta-analysis suggests that animals respond similarly to aircraft noise, as compared to other types of anthropogenic noise. Also, habitats that host breeding birds are generally closer to roadways than runways, therefore the actual received noise levels, and thus behavioral alterations, may be functionally similar between aircraft and traffic noise.

To date, there has been much research on how animals, especially birds, respond to anthropogenic noise. These studies have found altered habitat use, as well as altered communication, such as increased vocal behavior, decreased vocal behavior, and/or shifts in the timing of vocal behavior in noisy habitats. Some studies have also found that exposure to anthropogenic noise is associated with reduced reproductive success. Species-specific responses to noise (e.g., altered timing of vocalizations) may indicate that behavioral plasticity allows some individuals to avoid negative consequences of human-induced environmental change. However, noise-induced behavioral plasticity does not always improve communication and may also create a tradeoff between communication success and mate attraction.

The intermittent and relatively unpredictable nature of some sources of anthropogenic noise, like aircraft events, may increase the likelihood of negative effects on wildlife, in certain scenarios. Indeed, unpredictable traffic noise resulted in greater behavioral alterations during the breeding period in Greater sage-grouse (Centrocercus urophasianus), as compared to chronic noise like natural gas drilling rigs. Exposure to intermittent aircraft noise and sonic booms, a phenomenon caused by aircraft traveling at speeds greater than the speed of sound in air, has also resulted in startle responses, and altered parental behaviors, growth and reproduction in domestic animals. Therefore, the intermittent and relatively high amplitude noise associated with early morning flights (0500–0800 h) may be associated with negative impacts on individual and/or population health for birds breeding near airports or along flightpaths. These negative consequences may be driven by the overlap

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of early morning aircraft noise with dawn chorus, the period of time immediately surrounding dawn during which passerines display their highest rates of vocal behavior to attract mates and maintain territories. In the Northern hemisphere, this early dawn chorus is most pronounced during the spring breeding season, as early singing behavior avoids singing during optimal foraging times later in the day (when temperatures are warmer). Therefore, alteration to vocal behavior during dawn chorus may result in decreased pairing success or increased energetic demands.

Here, we investigated the impacts of aircraft events on vocal behavior in the wood thrush (Hylocichla mustelina) breeding in Sapsucker Woods Sanctuary (hereafter ‘SSW’), a bird sanctuary located approximately one mile from the Ithaca Tompkins International Airport (ITH), NY, USA. We used passive acoustic monitoring techniques to assess vocal behavior in wood thrush on days that differed in the number of aircraft events. Specifically, we calculated a ‘flight score’ for each 10-min period between 0500 and 0800 h (dawn chorus). This ‘flight score’ parameter was the sum of the number of flights that occurred immediately before and during each 10-min period from 0500 to 0800 h (see methods below). No birds were directly handled during this study. The wood thrush is a suitable species in which to study anthropogenic noise impacts, as they, like other passerines, rely on song to attract mates and maintain territories. Wood thrush vocalizations span a frequency range of approximately 2–9 kHz (Fig. 4A). Therefore, wood thrush can likely hear below 2 kHz, where much of the sound energy associated with aircraft events occurs.

We hypothesized a negative relationship between exposure to noise from aircraft events and wood thrush vocal behavior during dawn chorus, such that the number of wood thrush songs would decrease with greater exposure to aircraft events. We also predicted a negative relationship between the number of sites with wood thrush song (our proxy of the number of wood thrush) and aircraft events from 0550 to 0650 h, the peak of morning aircraft events at the Ithaca Tompkins Regional Airport. If our results show a negative relationship between aircraft events and vocal behavior, wood thrush may suffer reduced mate pairing success in habitats near airports or along flightpaths.

Results

Song rate. Our results showed a positive relationship between flight score and wood thrush song output when taking into account time of day and distance from the airport. Specifically, the number of songs at sites ‘close’ to the runway was higher during the peak of aircraft activity (0550–0650 h; Fig. 1A). The best-ranked model of wood thrush vocal behavior included a three-way interaction between ‘time’, ‘flight score’ and ‘distance group’ (Table 1). Also, the 95% CIs for the ‘flight score’ parameter, as well as the interaction effect between ‘time’ and ‘flight score’, do not overlap zero (Table 2). Unsurprisingly, there was a negative relationship between ‘time’ and wood thrush vocal behavior, as peaks in wood thrush song rate occur early in the morning (Fig. 1A–C).

Number of sites with song. The mean number of sites with wood thrush song during the peak of aircraft activity (our proxy of the number of wood thrush) did not differ based on flight score (Fig. 2). The null model was best-ranked and received approximately 75% of the model weight (Table 3).

Discussion

Our results showed a positive relationship between aircraft events and wood thrush vocal behavior during certain time periods, which was opposite our hypothesized negative impact of aircraft events on wood thrush song. This pattern was strongest at sites closest to the runway, with a 387% average increase in wood thrush vocal behavior over baseline when flight scores equaled 1 (Fig. 1D) and a 266% average increase over baseline when flight scores equaled 2 (Fig. 1E). These results are consistent with the findings of previous studies of noise pollution on song patterns in birds, as well as other taxa. Additionally, natural noise sources, such as waterfalls and torrents, have been found to increase song redundancy in chaffinches (Fringilla coelebs). However, it is worth noting that other studies have found animals, like humpback whales (Megaptera novaeangliae), to be less likely to produce vocalizations given anthropogenic noise. The altered vocal behavior found here may compensate for any acoustic masking of song during brief aircraft events. Alternatively, altered song may create a tradeoff between communication success and mate attraction.

The fact that our data also showed no relationship between aircraft events and the number of sites with wood thrush song during the peak of aircraft activity suggests that the increase in singing activity at sites closer to the airport is not explained by stimulating more individuals to sing when flights scores equal 1 or 2, as compared to baseline. Rather, aircraft events may stimulate increased song rate within individuals. Future research investigating if noise-exposed wood thrush also alter the pitch of their song, as has been found in urban great tits, and/or the complexity of their song, as has been found in red-winged blackbirds exposed to experimental noise, would be of great value. Based on our data set, the timescale at which altered vocal behavior may persist in wood thrush is also unclear. However, past studies have found that other passerines return to baseline song frequency and timing patterns relatively soon after experimental noise exposure. Additionally, American black ducks (Anas rubripes) and European seabass (Dicentrarchus labrax) have been found to behaviorally habituate to variable sound impulses. The intermittent and somewhat unpredictable nature of aircraft events at the Ithaca Tompkins International Airport (due to inconsistencies in flight schedules) may make the likelihood of habituation to aircraft noise relatively low for wood thrush in our study system.

Again, because habitat quality for wood thrush differs throughout SSW, regardless of distance from the airport, we cannot make general claims about the effects of aircraft noise on the distribution of wood thrush throughout SSW. Yet, our data suggest that forested sites that are only 450 m from the runway are still used by breeding wood thrush. This lack of avoidance of noise-exposed habitats has also been found in other passerines, such as serins. However, it is worth noting that other studies have found that birds avoid noise-exposed habitats.
both during breeding\cite{44} and migration\cite{45}. It is possible that aircraft noise levels above a certain threshold (which was not reached in this study) could result in reduced wood thrush vocal behavior and/or habitat use (i.e. a dose response). Dose–response models in killer whales (\textit{Orcinus orca}) show avoidance behavior when exposure to ship noise is greater than 150 dB re 1 μPa\cite{46}. Also, zebra finches (\textit{Taeniopygia guttata}) increase the amplitude of their song with increasing background noise levels, but cease to sing once background noise levels reached 80 dB\cite{47}.

\begin{table}
\centering
\begin{tabular}{|l|l|l|l|l|}
\hline
\textbf{Model} & \textbf{K} & \textbf{AICc} & \textbf{ΔAICc} & \textbf{Weight} \\
\hline
Time + FlightScore + Distance Group + Time*FlightScore + Distance Group + Time*Distance Group  & 15 & 5109.34 & 0 & 0.999 \\
Time + FlightScore + Distance Group + Time*Distance Group  & 9 & 5128.94 & 19.603 & <0.001 \\
Time + FlightScore + Distance Group + FlightScore*Distance Group  & 10 & 5129.34 & 20.000 & <0.001 \\
Time + Distance Group  & 8 & 5139.21 & 29.877 & <0.001 \\
Time  & 5 & 5156.73 & 47.393 & <0.001 \\
\hline
\end{tabular}
\caption{Model comparisons for the relationship between aircraft noise and wood thrush song number. *Date and site were included in all models as random effects.}
\end{table}
Finding a noise threshold at which behavioral changes, and perhaps negative fitness impacts, are expected in wood thrush, would be highly useful in this system and are an area of future focus.

The potential conservation implications for increased vocal behavior during time periods of increased air-craft noise are unclear. The wood thrush is a near-threatened species on the IUCN ‘red list’, and their population declines are thought to be related to reduced availability of forested breeding habitats48,49. Shifts in vocalizations by birds experiencing anthropogenic noise have been associated with positive, negative, or no effects on fitness25,50–52. Here, it is possible that the increased wood thrush song rate is associated with slightly increased energetic demands53,54 and/or time tradeoffs with other important behaviors, such as foraging55. Alternatively,
wood thrush may suffer no negative fitness consequences given increased vocal behavior, as the number of flights, and thus the time period during which vocalizations may be masked, is relatively low for the Ithaca Tompkins International Airport. Overall, the implications of increased vocalizations are unknown in this system. Although logically difficult, future work investigating the actual fitness costs associated with altered vocal strategies for wood thrush in noise-exposed habitats would be of high value, and may be applicable across noise type. Additionally, knowing the noise thresholds associated with fitness costs (if any) and how different behavioral strategies (i.e. changing the pattern of vocalizations) may allow individuals to evade these costs would be useful for establishing conservation policy in breeding habitats used by passerines, such as wood thrush.

Methods

Sound recordings. Recordings were taken at 30 sites in SSW (Ithaca, NY, USA; Lat 42.481673, Long −76.460053, Fig. 3A), using autonomous recording units (Swifts, Center for Conservation Bioacoustics, Cornell University) mounted to trees, approximately 2 m off the ground and spaced at least 250 m apart. Swifts recorded in a single channel (mono) from 1 to 19 May 2017, from 0500 to 0800 h at a 48 kHz sampling rate, with a 16-bit resolution. The sensitivity of the omnidirectional microphone was −44 dB re 1 V/Pa (± 2 dB) and the frequency response is flat (± 3 dB) in the relevant frequency range 10 Hz to 10,000 Hz. The gain was set to 33 dB and the clipping level of the analog to digital converter (ADC) is ± 0.9 V. Swift sites varied in their distance from the airport runway (456–1322 m), resulting in differential exposure to aircraft noise (Fig. 3B). Daily weather conditions were similar in Ithaca, NY, USA from 1 May to 19 May 2017 (i.e. no days in which rain functionally altered wood thrush song).

Sites were categorized as ‘close’ (450–750 m, n = 7 sites), ‘mid-distance’ (750–1050 m, n = 16 sites), or ‘far’ (1050–1350 m, n = 7 sites) from the Ithaca Tompkins International Airport runway for analysis. These three distance categories were chosen because they equally split the range of distances (456–1322 m) from the runway that existed in our data and created a parameter across which peak exposure to aircraft noise differed between groups (ex. 5 May 2017: ‘close’, 73.8 dBA ± 3.9 SD, ‘mid-distance’, 69.3 dBA ± 5.2 SD, ‘far’, 67.3 dBA ± 5.5 SD). Distance from the runway was measured from the GPS coordinates of each recording location (in decimal degrees, accurate to approximately ± 1 m) and the center of the only runway at Ithaca Tompkins International Airport. Additionally, habitat type varies throughout SSW, with larger, denser portions of forest located further from the airport (see Fig. S2). Because wood thrush prefer these larger, denser forests, and thus might be more likely to colonize sites further from the airport for reasons unrelated to aircraft noise, we focused on within distance-group variation for analysis of vocal behavior.

We used a band-limited energy detector in Raven Pro 2.0 (Center for Conservation Bioacoustics, Cornell Lab of Ornithology) to identify all flight times between 0500 and 0800 h on focal days (minimum frequency = 200 Hz, maximum frequency = 1000 Hz, minimum duration = 10 s, maximum duration = 45 s, minimum separation = 60 s, minimum occupancy = 30%, SNR = above 20 dB, block size = 180 s, hop size = 90 s, percentile = 20%). The detector output was then manually checked by researchers who listened to each auto-detection to ensure it was an aircraft event. The ambient noise levels (L10 dBA) in SSW were calculated from 0600 to 0700 h on a subset of days in May 2017 (see supplement). Based on these data, ambient noise levels in SSW differed during time periods associated with aircraft overflights (5 May: 69.1 L10 ± 7.2 SD, 9 May: 60.9 L10 ± 1.9 SD, 19 May: 65.8 L10 ± 4.8 SD), as compared to time periods without overflights (5 May: 58.5 L10 ± 2.5 SD, 9 May: 54.5 L10 ± 2.4 SD, 19 May: 58.2 L10 ± 3.6 SD, see supplement for more details). Given that an increase of 10 dBA is perceived as twice as loud in humans and passerines’ perceptions of sound is thought to be similar, but slightly less sensitive, than humans, our data suggest that aircraft events have a substantial impact on the soundscape for birds breeding in SSW (Fig. 3C and Supplementary Fig. S1). Also, aircraft events altered the acoustic signature of SSW for an average of 47.5 ± 27.4 ± (SD), similar to previous studies of aircraft noise profiles.

Sound analysis. We analyzed the number of wood thrush songs in a subset of recordings in May 2017 (n = 10 days, Mon–Fri only, May: 1, 2, 3, 4, 5, 8, 9, 15, 17, 18). These specific dates were chosen because they fell within the early breeding period for wood thrush in Ithaca, NY, USA, when singing for mate attraction and territory establishment is most common. These dates also varied in their flight schedule and time of first flight, thus allowing us to control for the potential impacts of time of day, time of first flight, and number of flights on wood thrush vocal behavior. On each focal day, we identified all sites with wood thrush song from 0500 to 0800 h and used data from these sites for our analyses.

To decrease the likelihood of duplicate song records from a given individual in our analysis, we eliminated sites with simultaneous song if the sites were less than 500 m apart (i.e. adjacent Swift units). Given that we do not have field data on the number of wood thrush territories present in SSW during May 2017, the presence of distinct wood thrush song across recording sites is our best estimate of the number of wood thrush singing in any given 10-min time period. It is important to acknowledge that wood thrush territory size ranges from 0.08 to 4 ha (approximately 800–40,000 m²). Therefore, we may have had multiple wood thrushes recorded by one unit, or multiple recording units within a single wood thrush territory. Nonetheless, this limitation should not impact our ability to identify relationships between aircraft events and vocal behavior, as we focused our analysis on changes in vocal behavior compared to a baseline level (flight score = 0; see below for details). Therefore, the spatial data found here can provide useful information on the relationship between aircraft events and wood thrush vocal behavior within SSW and across time.

We used Raven Pro 2.0 to visualize sound recordings and identify all sites with wood thrush song before, during, and after aircraft events between 0500 and 0800 h (Fig. 4B). Researchers trained in wood thrush song identification used template detectors in Raven Pro 2.0 (detection frequency range = 200 Hz and threshold = 0.6–0.8.
Figure 3. Maps and spectrograms associated with recording sites. (A) Aerial map of the Ithaca Tompkins International Airport and Sapsucker Woods (outlined in white, with recording locations marked by white points), (B) Distribution of peak amplitude levels (dBA) associated with an aircraft overflight across recording sites in Sapsucker Woods (greater received amplitude levels indicated in red, lower received amplitude levels indicated in blue), (C) Spectrogram (kHz vs. h) from 0500 to 0800 h at one recording site in SSW. Aircraft overflights indicated by arrows. Satellite images for A) and B) were generated from GoogleEarth (version 7.3.2, https://earth.google.com/web/search/Sapsucker+Woods,+Ithaca,+NY/).
depending on quality of recording, with loose matching and non-merged templates) to detect and annotate each individual wood thrush song (Fig. 4A). Detector outputs were also manually checked to ensure accuracy. In addition, we analyzed recordings to assess the relationship between aircraft events and the number of sites with wood thrush song (our proxy of the number of wood thrush). Specifically, we focused on the peak of morning aircraft activity at the Ithaca Tompkins International Airport, which occurs from 0550 to 0650 h. These data on the number of sites with wood thrush song allow us to contextualize any differences in song number. For example, if we found a negative relationship between aircraft events and wood thrush song number, but no difference in the number of sites with wood thrush song, our results would suggest that wood thrush are reducing song rate to some non-zero value, but not leaving their territories or completely ceasing song production during time periods affected by aircraft noise.

To assess whether sites had thrush territories present, we again used Raven Pro 2.0 to visualize sound recordings from 0550 to 0650 h, the peak of morning aircraft activity at our sites. Therefore, our data represented the presence or absence of wood thrush song in each 10-min time period at each site during May 2017. We then compared the number of sites with wood thrush song in each 10-min time period across flight scores (flight score = 0, 1, or 2; see paragraph on statistical analysis below). Similar to our previously described methods, we only included simultaneous song data if the sites were greater than 500 m apart to avoid duplicate bird song in our analysis. Unfortunately, there were no days in May 2017 in which zero flights went out from the Ithaca Tompkins International Airport between 0550 and 0650 h; therefore, we cannot gauge wood thrush vocal behavior in the complete absence of aircraft events.

To better understand the relationship between aircraft noise and wood thrush vocal behavior across time, we summed the number of songs per 10-min time period at each site, from 0500 to 0800 h. Wood thrush vocal behavior during a focal 10-min time period may be affected by flight(s) that occurred in that time period, as

Figure 4. Spectrograms associated with recoding sites. (A) Spectrogram (kHz vs. s) of four wood thrush songs, with each song boxed in white. (B) Spectrogram (kHz vs. s) of wood thrush songs before, during and after an aircraft event (start of aircraft event indicated by the white arrow).
well as any flights that occurred earlier in the morning. To identify the extent to which flight history affects wood thrush songs in any successive 10-min time period, we ran a partial least squares analysis which included a flight number parameter for each focal 10-min time period (e.g. FlightNumber\(_T\) + FlightNumber\(_{T-1}\) + FlightNumber\(_{T-2}\) … FlightNumber\(_{T-13}\), n = 14). The partial least squares analysis indicated that the error was lowest with the first two parameters (e.g. FlightNumber\(_T\), FlightNumber\(_{T-1}\), Fig. 5). Therefore, we created a ‘flight score’ parameter that was the sum of the number of flights that occurred in the focal 10-min time period, as well as the 10-min time period prior (i.e. if the focal time period was 0600–0610 h and there was one flight at 0552 and another flight at 0607, the flight score associated with the focal time period would be 2). We then ran generalized linear models, with the number of wood thrush songs per 10-min time period as the dependent variable. In each model in our model set, we included ‘site’ and ‘date’ as random effects, and ‘time’ as a fixed effect. We included ‘date’ in these analyses to account for the change in sunrise (and thus the expected time at which birds will start singing) and the reproductive stage of individuals. Additional models included a combination of the main effects and/or interaction effects for ‘distance group’ and ‘flight score’ parameters. Both ‘distance group’ and ‘flight score’ were included as fixed effects (see Table 2 for the full model set).

For our second analysis, data recorded the presence or absence of wood thrush song in each 10-min time period, across sites and days. Therefore, we ran generalized linear mixed models with binomial distributions, using the \textit{glmer} function in the \textit{glmm} package in R. All models in this model set included ‘date’ as a random effect. We also included models that accounted for both the main and/or interaction effects between ‘time’ and ‘flight score’ (see Table 1 for the full model set).

For the models associated with each analysis, we used the values of Akaike Information Criterion corrected for small sample sizes (AICc) and model weights for model comparisons\(^5\). These comparisons allowed us to evaluate the hypotheses that exposure to aircraft events was related to our dependent variables: if models including the measure of ‘flight score’ were ranked higher than models that did not include this term, our results suggest that aircraft events explain a considerable proportion of variation in our data. For the top-ranked models that included the ‘flight score’ parameter, we estimated the effect size (\(\beta\) parameter estimates) and 95% confidence intervals (CI) of our noise parameters through the \textit{summary} function in R. We assessed the importance of effect sizes based on whether the 95% CI overlapped zero (Table 2).

**Data availability**

Data will be made available on Dryad upon acceptance.

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**References**

1. Barber, J. R., Crooks, K. R. & Fristrup, K. M. The costs of chronic noise exposure for terrestrial organisms. \textit{Trends Ecol. Evol.} \textbf{25}, 180–189 (2010).
2. Buxton, R. T. \textit{et al.} Noise pollution is pervasive in US protected areas. \textit{Science} (80-) \textbf{356}, 531–533 (2017).
3. Manci, K. M., Gladwin, D. N., Villeta, R. & Cavendish, M. G. Effects of aircraft noise and sonic booms on domestic animals and wildlife: a literature synthesis (Engineering and Services Center U. S. Air Force, 1988).
4. Pott-Pollenske, M. et al. Airframe noise characteristics from flyover measurements and prediction. In 12th AIAA/CEAS Aeroacoustics Conference (27th AIAA Aeroacoustics Conference) 2567 (2006).
5. Khaldi, S. Reduction of commercial aircraft noise emission around airports. A new environmental challenge. Eur. Transp. Res. Rev. 1, 175–184 (2009).
6. Dooling, R. J. & Popper, A. N. The effects of highway noise on birds (The California Department of Transportation Division of Environmental Analysis, 2007).
7. Etzel, R. A. & Balk, S. J. Pediatric environmental health (American Academy of Pediatrics, Itasca, 2011).
8. Schomer, P. D. Growth function for human response to large-amplitude impulse noise. J. Acoust. Soc. Am. 64, 1627–1632 (1978).
9. Kunc, H. P. & Schmidt, R. The effects of anthropogenic noise on animals: a meta-analysis. Biol. Lett. 15, 20190649 (2019).
10. Shannon, G. et al. A synthesis of two decades of research documenting the effects of noise on wildlife. Biol. Rev. 91, 982–1005 (2016).
11. Slabbekoorn, H. et al. A noisy spring: the impact of globally rising underwater sound levels on fish. Trends Ecol. Evol. 25, 419–427 (2010).
12. Brown, A. L. Measuring the effect of aircraft noise on sea birds. Environ. Int. 16, 587–592 (1990).
13. McLaughlin, K. E. & Kunc, H. P. Experimentally increased noise levels change spatial and singing behaviour. Biol. Lett. 9, 20120771 (2013).
14. Injuan, A. S., Poon, L. Y. & Patricelli, G. L. Effects of experimental anthropogenic noise on avian settlement patterns and reproductive success. Behav. Ecol. 29, 1181–1189 (2018).
15. McClure, C. J. W., Ware, H. E., Carlisle, J., Kallenecker, G. & Barber, J. R. An experimental investigation into the effects of traffic noise on distributions of birds: avoiding the phantom road. Proc. R. Soc. London B Biol. Sci. 280, 20132290 (2013).
16. Kruger, D. J. D. & Du Preez, L. H. The effect of airplane noise on frogs: a case study on the Critically Endangered Pickersgill’s reed frog (Hyperolius pickersgilli). Ecol. Res. 31, 393–405 (2016).
17. Melcon, M. L. et al. Blue whales respond to anthropogenic noise. PLoS ONE 7, e32681 (2012).
18. Siervo, J., Schloesing, E., Pavón, I. & Gil, D. European blackbirds exposed to aircraft noise advance their chorus, modify their song and spend more time singing. Front. Ecol. Evol. 5, 68 (2017).
19. McCarthy, E. et al. Changes in spatial and temporal distribution and vocal behavior of Blainville’s beaked whales (Mesoplodon densirostris) during multiship exercises with mid-frequency sonar. Mar. Mammal Sci. 27, E206–E226 (2011).
20. Dominoni, D. M., Greif, S., Nemeth, E. & Brumm, H. Airport noise predicts song timing of European birds. Ecol. Evol. 6, 6151–6159 (2016).
21. Gil, D., Honarmand, M., Pascual, J., Pérez-Mena, E. & Macías, G. C. Birds living near airports advance their dawn chorus and reduce overlap with aircraft noise. Behav. Ecol. 26, 435–443 (2014).
22. Habib, L., Bayne, E. M. & Boutin, S. Chronic industrial noise affects pairing success and age structure of ovenbirds Seiurus aurocapilla. J. Appl. Ecol. 44, 176–184 (2007).
23. Halfwerk, W., Holleman, L. J. M., Lessells, C. K. & Slabbekoorn, H. Negative impact of traffic noise on avian reproductive success. J. Appl. Ecol. 48, 210–219 (2011).
24. Wollenden, A. D., Slabbekoorn, H., Kluk, K. & de Kort, S. R. Aircraft sound exposure leads to song frequency decline and elevated aggression in wild chiffchaffs. J. Anim. Ecol. 88, 1720–1731 (2019).
25. Halfwerk, W. et al. Low-frequency songs lose their potency in noisy urban conditions. Proc. Natl. Acad. Sci. 108, 14549–14554 (2011).
26. Bickley, J. L., Blackwood, D. & Patricelli, G. L. Experimental evidence for the effects of chronic anthropogenic noise on abundance of Greater Sage-Grouse at leks. Conserv. Biol. 26, 461–471 (2012).
27. Pepper, C. B., Nascarella, M. A. & Kendall, R. J. A review of the effects of aircraft noise on wildlife and humans, current control mechanisms, and the need for further study. Environ. Manag. 32, 418–432 (2003).
28. Staicer, C. A., Specter, D. A. & Horn, A. G. The dawn chorus and other diel patterns in acoustic signaling. In Ecology and evolution of acoustic communication in birds, 426–453 (1996).
29. Gil, D. & Llusia, D. The bird dawn chorus revisited. In Coding strategies in vertebrate acoustic communication 45–90 (Springer, Berlin, 2020).
30. Warren, P. S., Katti, M., Ermann, M. & Brael, A. Urban bioacoustics: It's not just noise. Anim. Behav. 71, 491–502 (2006).
31. Dooling, R. Avian hearing and the avoidance of wind turbines (University of Maryland, College Park, 2002).
32. Halfwerk, W., Both, C. & Slabbekoorn, H. Noise affects nest-box choice of 2 competing songbird species, but not their reproduction. Behav. Ecol. 27, 1592–1600 (2016).
33. Ware, H. E., McClure, C. J. W., Carlisle, J. D. & Barber, J. R. A phantom road experiment reveals traffic noise is an invisible source of habitat degradation. Proc. Natl. Acad. Sci. 112, 21015–21019 (2015).
34. Williams, K., Erbe, C., Ashe, E., Beerman, A. & Smith, J. Severity of killer whale behavioral responses to ship noise: A dose–response study. Mar. Pollut. Bull. 79, 254–260 (2014).
35. Cynx, J., Lewis, R., Tavel, B. & Tse, H. Amplitude regulation of vocalizations in noise by a songbird Taeniopygia guttata. Anim. Behav. 56, 107–113 (1998).
48. Rushing, C. S., Ryder, T. B. & Marra, P. P. Quantifying drivers of population dynamics for a migratory bird throughout the annual cycle. *Proc. R. Soc. B Biol. Sci.* **283**, 20152846 (2016).
49. Stanley, C. Q. *et al.* Connectivity of wood thrush breeding, wintering, and migration sites based on range-wide tracking. *Conserv. Biol.* **29**, 164–174 (2015).
50. Kleist, N. J., Guralnick, R. P., Cruz, A. & Francis, C. D. Anthropogenic noise weakens territorial response to intruder’s songs. *Ecosphere* **7**, e01239 (2016).
51. Ward, S., Speakman, J. R. & Slater, P. J. B. The energy cost of song in the canary, *Serinus canaria*. *Anim. Behav.* **66**, 893–902 (2003).
52. Nemeth, E. & Brumm, H. Birds and anthropogenic noise: are urban songs adaptive?. *Am. Nat.* **176**, 465–475 (2010).
53. Oberweger, K. & Goller, F. The metabolic cost of birdsong production. *J. Exp. Biol.* **204**, 3379–3388 (2001).
54. Ophir, A. G., Schrader, S. B. & Gilmore, J. F. Energetic cost of calling: general constraints and species-specific differences. *J. Evol. Biol.* **23**, 1564–1569 (2010).
55. Thomas, R. *et al.* The trade-off between singing and mass gain in a daytime-singing bird, the European robin. *Behaviour* **140**, 387–404 (2003).
56. Sheikh, P. A. & Uhl, C. Airplane noise: A pervasive disturbance in Pennsylvania Parks, USA. *J. Sound Vib.* https://doi.org/10.1016/j.jsv.2003.09.014 (2004).
57. Burnham, K. P. & Anderson, D. R. Multimodel inference understanding AIC and BIC in model selection. *Sociol. Methods Res.* **33**, 261–304 (2004).
58. Hurvich, C. M. & Tsai, C.-L. Regression and time series model selection in small samples. *Biometrika* **76**, 297–307 (1989).

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**Author contributions**

A.S.I. and H.K. conceived the ideas and designed the methodology, A.S.I. and E.D.L. collected the data, A.S.I. analyzed the data and led the writing of the manuscript. All authors contributed critically to the drafts and gave final approval for publication.

**Competing interests**

The authors declare no competing interests.

**Additional information**

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