Composition, impacts, and removal of liquid petroleum waste through bioremediation as an alternative clean-up technology: A review

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ABSTRACT

Exposure to hazardous wastes, especially petroleum wastes hydrocarbon (PWHCs), can damage human health and biological diversity. A huge amount of petroleum waste along with persistent organic pollutants is being generated during exploration and processing of crude oil. The dumping of petroleum waste hydrocarbons in an open pit contaminates the soil which can cause severe threats to human health and agro-geo-environmental ecosystem. The current study aimed to evaluate the mode of occurrence, composition, environmental, and health impacts of petroleum waste by using recent literature. The extracted results show that oil emulsion contains 48% oil, suspension 23%, settled emulsion 42%, and sludge emulsion 36%. The study discusses the possible biological techniques for rehabilitation of petroleum waste-contaminated areas. Several physical and chemical techniques are available for remediation of petroleum waste, but they are either costly or environmentally not feasible. Whereas, biological remediation namely, Bioremediation (Biostimulation and Bioaugmentation), Phytoremediation (Phytodegradation, Rhizoremediation, Phytovolatilization, and Rhizo-filtration) is a cheap and environmentally friendly way to remove petroleum waste hydrocarbons from contaminated soil and water. Some important enzymes (i.e., peroxidase, nitrilase, nitroreductase, phosphatase) and plant species i.e., Acacia and Chloris species are prominent methods to remediate the PWHCs. The knowledge assembled in this review is expected to create new doors for researchers to develop more efficient techniques to control the harmful impacts of PWHCs on the environment and health.

1. Introduction

Hydrocarbon and its wastes are the aggregations of various materials, including crude oil, emulsion, solids, and slurry. The exploration, processing, and transportation of hydrocarbon and petroleum wastes may damage the surrounding environment, aquatic, and terrestrial biota as well as human health (Shahzad et al., 2020; Karam and Al-Wazzan, 2021). The most fragile contents of the environment namely, air, soil, and water are essential factors for the biogeochemical ecosystem and are exposed to hazardous nature of hydrocarbon (Nwaichi et al., 2014; Turner et al., 2021). Soil contamination occurs due to sporadic discharge and dumping of hydrocarbon and petroleum wastes. Thus, the petroleum wastes that affected soil do not support flora and flora and remained unfertile over decades before being remediated (Khan et al., 2018). Similarly, the overexploitation of petroleum resources also contaminates the water resources, especially small agricultural channels, rivers, and streams located nearby the oil rigs and processing units (Balogh and Watson, 2020). Due to lower density, oily wastes can float over the water surface and thus affect photosynthetic plants/organisms in water bodies and soil (because of settling on the surface). Whereas denser petroleum waste penetrates downward; causing groundwater contamination and reducing the soil porosity (Sayed et al., 2021). The contaminated water...
with hydrocarbon and persistent organic wastes drastically affects human health and other living organisms (Ahmed and Fakhruddin, 2018). Some petroleum waste is volatile and affects the air quality and causes other types of air pollution i.e., smog, which needs to formulate anti-air pollution strategies (Chen et al., 2020). The availability of hydrocarbon and petroleum wastes in air may also cause nose and throat allergies, vomiting and significantly affect the lungs of humans and other living organisms (Müller and Sedláčková, 2003). Exposure to a high concentration of hydrocarbon may cause malignant and non-malignant lungs disorder (Connellan, 2017). It is documented officially, that any type of exportation, processing, preparation, or operation activities that generate pollutants or contaminate the surrounding environment may need to be remediated or reduce the pollutant below the permissible level (GoP, 1997; Sattar et al., 2021).

More than 100 thousand sites all over the world are categorized as hazardous for living organisms, particularly humans. Most of these are mainly from the disposal of untreated petroleum hydrocarbons (Bujang et al., 2013). Complete remediation of such sites over short period has remained a problem, mostly because of cost involved in the disposal and processing. Other possibilities include lack of best remediation technique that can degrade the diverse nature of petroleum waste hydrocarbons over a short period (Jasmine and Mukherji, 2014). The spilling and open dumping of liquid oily wastes in the surrounding environment and their exploration and identification is one of the major problems. The purpose of this study was: 1) To explore the source and composition of petroleum waste in contaminated soils at dumping sites. 2) To evaluate the impacts of petroleum waste on human health, soil microbiota, land degradation, and aquatic life. 3) To discuss the possible available biological techniques for rehabilitation of petroleum waste-contaminated soils. The effective petroleum waste remediation action plan includes; the collection of petroleum waste from the source, analyzing, processing, and remediation of petroleum hydrocarbons through different physical, chemical, and biological methods (Figure 1).

2. Mode of occurrence of petroleum hydrocarbons

It was well reported that hydrocarbon is comprised of various hazardous materials depending upon the source, organic materials, geological environment, and decomposition rate (Ahmadun et al., 2009; Waples, 2013; Srivastava et al., 2019). This is more likely because petroleum is derived from residual organic matter or biomass of biogenic origin including terrestrial plants-C4 and marine algae-C3 or phytoplankton (Welte and Tissot, 1984). Moreover, marine algae have high content of organic materials, especially nitrogenous and lipid compounds, however, the terrestrial plants have high compositions of lignin, cellulose, and other carbon-containing materials (Cloern et al., 2002; Mannino and Rodger, 2004). The structure and behavior of hydrocarbon derived from algal or terrestrial plants differ from one another and produce various types of pollutants during the exploration. The variation in petroleum compounds is because of their decomposition environment, temperature, biomarkers, maturity level, and presence/availability of organic materials (Thangalazhy et al., 2012). The maturity biomarker is mostly associated with polysaccharides (i.e., aromatic hydrocarbon, steroids, hopanoids, and porphyrins), and terpenes. However, the non-biomarker maturity level is commonly attributed to aldehyde, aromatic, and isoprenoids (Thangalazhy et al., 2012). The presence and availability of maturity biomarkers and non-biomarker are attributed to organic materials derived from terrestrial and marine plants. During Ordovician era, the algal growth (C3-plants) was dominant, where the hydrocarbon was abundant in n-alkanes with high content of odd n-alkane while there was less concentration of n-dodecane (Hatch et al., 1987; Colcord et al., 2019). However, the aromatic hydrocarbon i.e., phenanthrene and naphthalene were abundant in oil derived from both marine and terrestrial plant materials (Table 1).

Asif (2010) documented that the biogenic derived oil in Kohat and Potwar Plateau originated from marine/terrestrial materials, which are categorized as Kerogen type II and III. In this region, the crude oil was dominated by saturated hydrocarbons, while other organic compounds i.e., aromatic hydrocarbon were available in minor amounts. The saturated hydrocarbons in crude oil are derived from organic matter, while aromatic hydrocarbons in crude oil are originated from olefinic and naphthenic compounds associated with biogenic decay of organic materials (Nissenbaum et al., 1985; Roushdy et al., 2010; Waples, 2013; Wilkes et al., 2020).

![Figure 1. Overall summarization and action plan of the PWHCs from origination to remediation.](image-url)
3. Petroleum waste composition

It is well understood that during exploration of crude oil, a huge amount of petroleum waste is generated, which shows the characteristics and composition of source rocks (Hu et al., 2013; Ozdemir et al., 2020). The chemical characteristic of petroleum is always similar to crude oil and somewhat to the source rocks which they are derived from. However, the major portion of petroleum is crude oil, while in trace amounts, emulsion, water, soil, and clay are available (Table 2). The generation of petroleum waste during exploration and its dumping in open pits may cause hazardous impacts on humans. Similarly, petroleum wastes may cause carcinogenic impacts because of the presence of carcinogenic chemicals (Benzene, anthracene, Cr⁶⁺, Pyrene, As) (Mansur et al., 2015). The composition of wastes in petroleum may differ from area to area and vary from region to region depending on the biogenic and geologic environment (Ozdemir et al., 2020). Singh et al. (2017) documented that the world petroleum reservoir is dominated by aromatic and aliphatic hydrocarbon. The aromatic and aliphatic hydrocarbon dominated with 75–85% fractions (i.e., gasoline, kerosene, naphtha, diesel, crude oil, and bitumen), while sulfur, nitrogen, and oxygen are present in minor/low concentrations (De Junet et al., 2013). The aforementioned data revealed that aliphatic hydrocarbon, namely paraffin, alkanes, alkenes, and aromatic hydrocarbon (anthracene, pyrene, naphthalene, and chrysene) are also dominant in liquid petroleum wastes, which may cause serious environmental pollution and disturbance in an ecosystem (Connell et al., 1980; De Junet et al., 2013). The environmental pollution caused by petroleum exploration, and processing needs to be remediated by using advanced biological techniques to limit the environmental consequences (Serafim et al., 2008).

| Biomarker                  | C Range | Indication                                                                 |
|---------------------------|---------|----------------------------------------------------------------------------|
| n-Alkanes, CPI’1          | C1–C31  | Marine, lacustrine algal source, C25, C17, C13 dominant                    |
| Acyclic isoprenoids       |         |                                                                           |
| Head to tail Pristane     | C29     | Chlorophyll, o- tocopherol, oxic, sub-oxic environments                    |
| Phytane                   | C29, C30, C34 | Terrestrial plant wax source C29, C30, C34, dominant            |
| Head-to-Head Botryococcene|         |                                                                           |
| Sesquiterpenoids,         | C35     | Terrestrial plants                                                        |
| Cadalene, Eudesmane       |         |                                                                           |
| Diterpenoids,             | C39     | High plant resins                                                         |
| Afralane, pimarane,       |         |                                                                           |
| Kaurane, retene.          |         |                                                                           |
| Tricyclic terpenes        | C19–C45 | Diagenetic products of bacterial and algal cell-wall lipids               |
| Tetracyclic Terpenes      | C39     | Degradation of pentacyclic triterpenoids                                  |
| Hopanes                   | C25–C40 | Bacteria                                                                  |
| Norhopanes                | C27–C40 | Anoxic marine                                                             |
| 2- and 3- methyl hopanes  | C23–C38 | Carbonate rocks                                                           |
| Benzo[b]indenopentanes    | C26–C56 | Carbonate environments                                                    |
| Hexahydrobenzoindenopentanes| C22–C56| Anoxic, carbonate asphaltenes                                              |
| Gammacerane               | C35     | Hypersaline environments                                                  |
| Oleaneanes, Lupanes       | C35     | Late cretaceous and tertiary flowering plants                              |
| Bicadinane                | C35     | Gymnosperm tree resin                                                     |
| β-Carotane                | C19–C23 | Arid, hypersaline                                                         |
| Steranes                  | C28–C30 | Eukaryote organisms, plants, and animals                                   |
| 24-n-Propylsteranes       | C28–C30 | Restricted to marine sediments                                            |
| 4-Methylsteranes          | C20     | Marine and lacustrine dinoflagellates                                      |
| Dinosteranes              | C26–C31 | Marine, Triassic, or Younger                                              |

3. Petroleum waste composition

| Region          | Type       | Appearance |
|-----------------|------------|------------|
| USA             | Slop oil   | Emulsion   |
|                 | Sludge     | Suspension |
| Libya           | Crude oil tank bottom sludge | Emulsion |
| World           | Petroleum waste | Emulsion |
| India           | Oily sludge | Emulsion   |
| Pakistan        | Crude oil  | Emulsion   |

| Region           | Type       | Appearance       | Oil | Water | Solids | References                   |
|------------------|------------|------------------|-----|-------|--------|-------------------------------|
| USA              | Slop oil   | Emulsion         | 48  | 40    | 12     | (Bider and Hunt, 1982)        |
|                  | Sludge     |                  | 23  | 53    | 24     |                               |
| Libya            | Crude oil tank bottom sludge | Emulsion  | 42.08 | 2.9   | 55.02  | (Mansur et al., 2015)        |
| World            | Petroleum waste | Emulsion  | 30–50 | 30–50 | 10–12 | (Hu et al., 2013)            |
| India            | Oily sludge | Emulsion         | 35.59 | 55.13 | 9.246  | (Kumar and Raj Mohan, 2013)   |
| Pakistan         | Crude oil  | Emulsion         | 0.2–2.3 | 0–3.6 |        | (OGDCL, 2017)                |

Table 1. An indication of the carbon source for the C range present in Petroleum (adapted from Welte and Tissot (1984)).

Table 2. The mode of occurrence and composition of Petroleum waste hydrocarbon around the world.
4. Impacts of petroleum waste contamination

4.1. Effect on human health

Several health issues like respiratory system disorders, miscarriages and infertility in women, birth defects, skin rashes, and childhood leukemia are associated with petroleum hydrocarbon contamination. Besides, there are high chances of carcinogenic, mutagenic, and teratogenic disorders, in most cases, it can damage deoxyribonucleic acid (Rengarajan et al., 2015). Moreover, petroleum contamination in the surrounding environment is a significant source of concern with a detrimental impact on human health. Such compounds enter the human body through different routes, i.e., skin, inhalation, dermal, and ingestion (Figure 2). The toxicity of petroleum-waste hydrocarbon is mainly dependent on the concentration and duration of exposure (Rengarajan et al., 2015; Murawski et al., 2021). For instance, even short-term exposure to an elevated level of some volatile petroleum hydrocarbons (such as toluene) is toxic to the central nervous system and can cause dizziness and headache. In addition, some adverse health symptoms like respiratory problems, skin and eyes irritation, and headache are reported in the workers exposed to petroleum hydrocarbons for long term (Levy and Nassett, 2011). International Program on Chemical Safety (IPCS) 2000 suggested that some chronic symptoms like vomiting, diarrhea, abdominal pain, nausea, confusion, and headache can be experienced by a person when exposed shortly to organic pollutants (WHO, 1984).

4.2. Effects of petroleum hydrocarbons on soil microbiota

Petroleum waste hydrocarbons contaminated soils may lose their functions i.e., soil fertility, porosity, regulation of water supply, degradation of decayed material, and providing habitat to microorganisms. Contaminated soils may lose or threaten several microbial species, which will ultimately affect the biodiversity of soil microorganisms (Maflana et al., 2021). Ren et al. (2015) explored the effect of pyrene pollution on bacterial population.

| Source                                    | Bacterial Species                                                                 | References                  |
|-------------------------------------------|-----------------------------------------------------------------------------------|----------------------------|
| Petroleum Refinery Effluents (PAH)        | Pseudomonas sp.                                                                   | (Liu et al., 2021)          |
| Petroleum Hydrocarbon                     | Pseudomonas, Stenotrophomonas, Achromobacter, Mesorhizobium, and Brucella (genera's) | (Zhang et al., 2021)        |
| Oily sludge                               | Pseudomonas aeruginosa                                                            | (Varjani et al., 2020)      |
| Oily sludge                               | Bacillus cereus, Bacillus altitudinis, Commonmonas (Delftia sp.), and Stenotrophomonas maltophilia | (Shahzad et al., 2020)      |
| Oil refinery sludge                       | Bacillus, Coprothermobacter, Rhodobacter, Pseudomonas, chromobacter, Desulfotobacter, Desulfosporosinus, Methanobacterium, Methanoseta, | (Roy et al., 2018)          |
| Oily sludge                               | Gordonia alkaliphila and Gordonia paraffinivorans                                | (Qi et al., 2017)           |
| Oil contaminated soils (TPH)              | Pseudomonas stutzeri BP10 and Aspergillus niger PS9                               | (Kumari et al., 2016)       |
| Total Petroleum Hydrocarbons              | Pseudomonas sp.                                                                   | (Xu et al., 2016)           |
| Crude oil                                 | Pseudomonas aeruginosa                                                            | (Varjani et al., 2015)      |
| Petroleum hydrocarbons                    | Proteobacteria, Actinobacteria and Firmicutes (Phyla)                             | (Fuentes et al., 2014)      |
| Oily Sludge                               | Bacillus, Burkholderia, Paenbacillus, Pseudomonas, Bacillus, Stenotrophomonas Enterobacteria Pseudomonas, Bacillus, Pandoraeaand Kocuria | (Roy et al., 2014)          |
| Crude Oil                                 | Acinetobacter baumannii                                                           | (Kim et al., 2009)          |
| Petroleum contaminated soils (TEM, PAH, alkanes) | Pseudomonas sp. BS2201, BS2203 and Brevibacillus sp. BS2202 (Nitrate reducing bacterial strains) | (Gritschenkov et al., 2000) |
Their results depicted that bacterial population and diversity in pyrene-polluted soils were adversely affected. Moreover, the growth of several species of microorganism (i.e., belonging to phyla Chloroflexi, Alphaproteobacteria, Actinobacteria, Deltaproteobacteria, and Crenarchaeotae) was potentially decreased (Ren et al., 2015) (Figure 2, Table 3).

### 4.3. Land degradation

Quite numerous incidents happened in the world because of flushing/dumping of petroleum wastes in an open environment. However, the release of PWHCs may also occur because of accidental leakage of the storage tank and pipelines. Due to this, several fertile agricultural lands were converted to unfertile and degraded soil fertility (Zabchuk and Olsson, 2017). In 2003, an oil spillage accident occurred in Karachi, Pakistan, where about 15,000 tons of oil spilled into the marine environment from a crude oil tanker. The land was badly affected, which ultimately affected flora and fauna. Moreover, marine and wildlife were also affected. During transportation, oil spilled accidents may affect agricultural land nearby the highways. Similarly, several acres of agricultural land are annually getting contaminated with petroleum waste, nearby the waste pit because of leakage from waste pit (Siddiqui et al., 2016; Sattar et al., 2021). Therefore, it is necessary to control the further loss of agricultural land and need reclaimation of these land at any cost. The natural degradation process and rehabilitating of contaminated land is a slower process than the contamination of useable land (Shahzad et al., 2020). For the reclamation and rehabilitation of such sites, it is mandatory to either avoid further degradation of contaminated lands or to adopt some suitable remediation technology to remove petroleum waste from contaminated soils (Figure 2).

### 4.4. Effects on aquatic ecosystem

Petroleum seepage in surface water causes an adverse impact on the aquatic ecosystem (Figure 2). It not only causes marine pollution but also contaminates the fishing areas and drinking water sources (Charles et al., 2021). Besides, groundwater gets contaminated with petroleum hydrocarbons when oil percolates into soil and reaches the groundwater aquifers. Several scientific studies explored the negative impacts of petroleum hydrocarbons on aquatic biota (Dupuis and Francisco, 2015; Luke and Odokuma, 2021). The Immuno-toxic effects of petroleum hydrocarbons on aquatic biota (Dupuis and Francisco, 2015; Siddiqui et al., 2016; Sattar et al., 2021). Therefore, it is necessary to control the further loss of agricultural land and need reclaimation of these land at any cost. The natural degradation process and rehabilitating of contaminated land is a slower process than the contamination of useable land (Shahzad et al., 2020). For the reclamation and rehabilitation of such sites, it is mandatory to either avoid further degradation of contaminated lands or to adopt some suitable remediation technology to remove petroleum waste from contaminated soils (Figure 2).

### 4.5. Removal of petroleum waste

The presence of toxic and complex compounds in petroleum waste contaminated soils makes it difficult to completely remove PWHCs through the natural process of degradation. Up till now, several methods have been established to reclaim contaminated sites, including physical, chemical, and biological processes (Abdullah et al., 2020; Ossai et al., 2020; Sayed et al., 2021).

The selection of a suitable method for rehabilitation of contaminated sites is mainly dependent on composition and concentration of petroleum hydrocarbons. Among the above-mentioned techniques, physical and chemical processes are more efficient but they are either costly or ecotoxic. Whereas, biological processes are economical and environmentally friendly but take more time to reduce contamination from the soil surface.

### 5. Bioremediation

Bioremediation is an effective natural method for the degradation of pollutants (like PWHCs) using microorganisms. In microbial remediation, microorganisms attach themselves to petroleum hydrocarbons and degrade them by using them as a source of food. The process of bioremediation can be enhanced by adding microbial culture (bio-degraders) and fertilizers. However, the rate of biodegradation of different compounds is different (Namukuye, 2021; Zhen et al., 2021).

Microbes in the soil perform various activities i.e., chemical, and biological processes to degrade PWHCs effectively. However, microbes can play an essential role in the biogeochemical cycles (Sridhar et al., 2021). The microbial activities may be catabolic or metabolic using an enzyme to catalyze organic wastes in soil (Fierer et al., 2021). Microbial population and their communities are quite variable in different ecosystems. The microbial degradation of petroleum wastes may also depend on climatic conditions, biotic and abiotic factors, especially the availability of nitrogen, carbon, and oxygen.

It was reported that petroleum wastes especially in agricultural soil affect the soil, food quality, and yield. However, petroleum wastes in the soil can be degraded by microbes naturally, but the process will be very slow (Sengupta and Pal, 2021). The petroleum materials are less dense and may accumulate on the soil surface. Further, it can block the soil porosity and permeability which reduce microbial activity and plant growth (Hewelke et al., 2018). By blocking soil porosity and permeability, an anxious condition is created where oxygen availability for microbes in soil reduces which stops the microbial activity (Das and Kazy, 2014). Thus, bioremediation (by inoculating biodegrading microbes) can trigger the degradation of PWC’s contaminated soils. However, consortium of *Bacillus altitudinis* (KF859970), *Bacillus cereus* (KR232400), and *Comamonas* (KF859971) can trigger oil waste degradation, especially alkane group compounds (Shahzad et al., 2020) (Table 3). They further reported that the lag phase was minimized to 03 days with consortiums and 14 days without using consortiums. However, Maddela et al. (2017) reported that 87.5% of petroleum wastes were degraded using a synthetic microbial consortium of *Bacillus cereus*, *Bacillus thuringiensis* along with *Geomyces sp*., and *Geomyces pannorum HR* strain.

Although bioremediation is thought to be an effective strategy for overcoming all the obstacles to remediating petroleum hydrocarbon successfully, like other techniques of remediation. However, it also has

**Table 4. Various studies report the effectiveness of Bioaugmentation and Biostimulation of Petroleum hydrocarbons.**

| Contamination | Location | Technique applied | Reference |
|---------------|----------|------------------|-----------|
| 1 | Light crude oil contaminating a sandy beach | Marine oil wastes at sand beach, Delaware, USA | Bioaugmentation and biostimulation (Venosa et al., 1996) |
| 2 | Crude oil and heavy fuel oil | Sea Empress oil spill, UK | Biostimulation (Swannell et al., 1999) |
| 3 | Heavy crude oil | Japan | Bioaugmentation (Tsuitumi et al., 2000) |
| 4 | Crude oil degradation | The Grande Terre, Kerguelen islands | Biostimulation (Tellier et al., 2002) |
| 5 | Petroleum Hydrocarbons | Crude oil spill site in Liaohose Oilfield, Liaoning Province, China. | Bioaugmentation (Xu and Lu, 2010) |
| 6 | heavy crude oil/total petroleum hydrocarbons | Xingang port of Dalian City in China | Bioaugmentation (Cai et al., 2016) |
| 7 | Crude oil | the hyper-arid region in Israel | Biostimulation and Bioaugmentation (Banet et al., 2021) |
some limitations, such as inability to completely remediate and remove PWHCs from contaminated soils. Due to persistent nature of some compounds present in PWHCs contaminated soils, some hydrocarbons can be degraded completely; some resist the natural process of degradation, whereas some compounds are non-biodegradable (Artham and Doble, 2008; Patel et al., 2011). Similarly, saturated subsurface remediation (in anoxic conditions) may not be achieved effectively through this method. Additionally, this method cannot prevent groundwater contamination from PWHCs. Despite this, bioremediation with its limitations is widely accepted and applied. However, the improvement in bioremediation with genetically modified microorganisms, plants, and microorganisms producing surfactants is gaining slight attention among researchers.

6. Techniques of bioremediation

6.1. Biostimulation

Biostimulation provides optimized ecological conditions for bioremediation process in contaminated sites. This process is adopted when there is a need to speed up the natural degradation of contaminated sites (Zhou et al., 2019). As discussed above, when hydrocarbon-degrading microorganisms are injected into contaminated soils, PWHCs breakdown takes place. However, in contaminated soils, the degradation of PWHCs with nutrients improves up to the maximum level (Table 4). When nutrients are added to PWHCs' contaminated soils, however, more degradation can be achieved. The provision of oxygen, nutrients, and electron acceptors to bacteria (engaged in bioremediation) stimulates their development and activities. For this purpose, subsurface biostimulation systems are developed, and additives are injected through injection wells. However, some inconsistent evidence was also found in the literature regarding the PWHCs' degradation with nutrient and microbial consortium (Table 4). It was reported that consortiums along with nutrients had no observed impacts on petroleum waste hydrocarbon degradation (Siddiqui et al., 2016). However, a consortium along with nutrients triggered the degradation of PWHCs in contaminated soil effectively (Wu et al., 2017).

6.2. Bioaugmentation

Petroleum waste can also be remediated through bioaugmentation effectively. In this process, genetically engineered or indigenous microbes are added to the petroleum waste contaminated soils to degrade petroleum waste hydrocarbons (Cunningham and Philp, 2000). Bioaugmentation is an effective tool to enhance the degradation rate of hydrocarbons by increasing the total population of indigenous microbes (Wu et al., 2017). Bioaugmentation can enhance the rate of degradation by introducing non-indigenous microorganisms to contaminated sites. For the degradation of pollutants like petroleum hydrocarbons, bioaugmentation is encouraged. This usually involves modifying the pH of contaminated site, adding limiting nutrients to obtain an optimal C:N:P ratio, and increasing soil moisture. However, the bacteria that degrade petroleum hydrocarbons are exposed to hazardous oil spills, they attempt to release a suitable amount of biosurfactant which convert all type of hydrocarbon to some other substance, as a result minimizes the harmful impacts of compound (Table 4).

6.3. Land farming

Land-farming is an efficient tool used to degrade the PWHCs' contaminated soil. The contaminated soils are rotated periodically for aeration and mixed with soil nutrients to boost up enzymatic actions of microbes to increase the natural degradation process of petroleum hydrocarbons (Besalatpour et al., 2011; Wang et al., 2016). The land farming technique is usually preferred over landfilling and incineration as it involves less energy consumption, and a lower risk of water pollution (both surface and groundwater) due to the immobility of hydrocarbons through the soil, regulatory compliance, and compatibility with soil properties and climatic conditions (Besalatpour et al., 2011).

7. Phytoremediation

The remediation of oil wastes through microorganisms was further improved by the addition of plants. The microbial remediation of petroleum waste is considered bioremediation, while the oil waste remediation through plants is called phytoremediation. Thus both methods, either phytoremediation or bioremediation can reduce the petroleum waste in contaminated soil (Aliku et al., 2021). Phytoremediation is an environmentally friendly and economic tool to degrade and detoxify petroleum waste hydrocarbons. In this process, green plants are used to remove, disintegrate, stabilize, mineralize, or degrade the toxic pollutants from soil and water (Nero, 2021). Several plant species (i.e., Mirabilis Jalapa L, Sebastania commersoniana, Festuca arundinacea, Panicum virgatum) are reported as potential phytoremediators of petroleum hydrocarbons (Yavari et al., 2015) (Table 3). Such plants show good tolerance...
7.1. Phytodegradation

Phytodegradation is the process of degradation of contaminants using plant enzymes. However, organic substances can be degraded via phytodegradation either inside the plant or in rhizosphere. Solvents in the groundwater, petroleum, and aromatic compounds in soils are examples of such contaminants that can be removed from the environment using this technology. Some reported plants utilizing enzymes responsible for the degradation of organic contaminants are dehalogenase, peroxidase, nitrilase, nitrreductase, and phosphatase (Yadav et al., 2018) (Figure 3).

7.2. Rhizoremediation

Rhizoremediation is the most efficient way to degrade petroleum hydrocarbons (Alotaibi et al., 2021). Rhizoremediation involves both bioremediation and phytoremediation processes to degrade oily wastes in the surrounding environment. Rhizospheric soils provide a favorable environment for the growth and multiplication of petroleum hydrocarbons degrading microbes. The cumulative impact of secretion of root exudates (i.e., amino acids, organic acids, and sugars) and the presence of petroleum hydrocarbons degrading microbes enhance the rate of degradation. Hoang et al. (2021) investigated the rhizoremediation potential of Acacia pyrifolia, Banksia seminuda, Hakea prostrata, Triodia wiseana, Acacia inaequatoriae, Hardenbergia Violacea, Acacia stellaticeps, and Chloris truncata reduced the extent of total petroleum wastes in contaminated soil (Table 3). Furthermore, with the addition of Hakea prostrata, Banksia seminuda, and Chloris truncata, plants were efficient to degrade TPHs (total petroleum hydrocarbons) in diesel contaminated soils (Figure 3).

7.3. Phyto-volatilization

Phytovolatilization of petroleum hydrocarbons takes place when the pollutants (more specifically the VOC) in soil are uptaken by plant roots and evaporated through leaves by the process of evapotranspiration. The contaminants are converted to a gaseous state and thus released. This process is mostly applied in marshy environment or it needs to construct wetlands (Herath and Vithanage, 2015) (Figure 3). The overall graphical summary and process of remediation were given in Figure 1.

8. Factors affecting the biological degradation of petroleum waste hydrocarbons

The major factors, namely soil pH, temperature, viscosity, and oxygen can greatly affect the biological degradation of PWCs. The accomplishment of bioremediation is dependent on the presence of suitable biodegrading species of microbes, adaptation, composition of pollutants, and nutrient availability (Cameotra and Makkar, 2010). Besides above-mentioned factors, the efficiency of bioremediation further depends upon soil pH, temperature, oxygen supply, salinity, and viscosity of pollutants (Margesin and Schinner, 2001; Sharma, 2020).

The oily waste degradation efficiency of microorganisms also depends upon pH and the presence of calcite and dolomite. Literature shows that most PAH (polyaromatic hydrocarbons) degrading microbes perform effectively at neutral pH, whereas, fungi are reported to be tolerant in an acidic environment. Literature shows maximum degradation of PWHCs occurs at a pH range of 6–9 (Al-Daher et al., 1998; Sharma, 2020).

Temperature plays a significant role in the biodegradation of petroleum waste hydrocarbons. Literature shows that the occurrence, growth, and function of microorganisms in a medium are highly dependent on temperature. Eltoukhy et al. (2020); Gomez et al. (2021) reported when...
temperature is dropped to a level below the optimum growth temperature; the microorganisms responsible for the degradation are no more available. Moreover, literature shows that viscosity of petroleum wastes increases and volatility decreases at low temperatures, which ultimately slows down the degradation rate of petroleum waste hydrocarbons. Whereas, by increasing temperature, the solubility of petroleum waste hydrocarbons increases and viscosity decreases, so the solid phase hydrocarbons are converted to liquid phase. The maximum degradation occurs at a temperature range of 30–40 °C (Das and Tiwari, 2018).

It is well accepted that distribution pattern of petroleum waste hydrocarbons is different in the terrestrial environment; mainly depends upon (a) viscosity of PWHCs and (b) texture of the soil they are spilled on (Angeli and Hewitt, 2000). Highly viscous oil will be distributed horizontally on the land surface. Whereas the less viscous will penetrate vertically, their downward movement will depend upon the nature of soil they are spilled over; whether it’s sandy, silty, or clay soil. Moreover, viscosity of petroleum waste may also play a significant role in the plant’s growth. The viscosity index of ~35 shows lighter oil, which has a high potential of infiltration (regardless of the specific soil type) and groundwater contamination (Fingas and Brown, 2018; Fu et al., 2021; Jahromi et al., 2021).

The lighter petroleum waste blocks the soil porosity and creates anoxic environments in subsurface soil. The leaching of light petroleum waste in subsurface soil may also affect moisture content, gaseous and ion exchange capacity in subsurface soils. An anoxic environment and deficient moisture contents in subsurface soil affect the microbial community and nutrient availability (Patel et al., 2021; Wang et al., 2022) (Figure 4).

It is well documented that bioremediation accelerates the degradation of PWHCs under aerobic conditions (at the surface). Under anoxic conditions, the indigenous hydrocarbon-degrading microbial population decreases in number and diversity (Zhao et al., 2018; Fenibo, 2021). For instance, Roy et al. (2018) and Mai et al. (2021) conducted a bioremediation study on petroleum hydrocarbons, and their study concluded that bioremediation (under aerobic conditions) can effectively degrade petroleum wastes up to 15 m. However, at a higher depth, the degradation rate of petroleum wastes is minimized. Additionally, the effective degradation of PWHCs in soil is still under consideration. The researcher, therefore suggests, applying the soil flushing technique for the removal of PWHCs from subsurface soils rather than biostimulation by injecting air. However, installation of an injection well for oxygen supply is costly and may not be suggested for oily contaminated soil especially contaminated soil with PWHCs.

9. Conclusion

Petroleum waste released from oil industry/refinery is a global environmental issue. Petroleum waste hydrocarbons are carcinogenic, immune-toxic, mutagenic, and teratogenic in nature and may cause potentially harmful impacts on humans and other living organisms. Such compounds enter the human body through different routes, i.e., skin, inhalation, and ingestion. During last few decades, the environmental degradation of PWHCs has gained huge attention. The major contents in liquid petroleum wastes were water 55.2%, light hydrocarbon 23.2%, waxes 10.5%, clay 9%, and asphaltene 1.9%. Microorganisms are known for bio-transformation and bio-degradation of different types of environmental contaminants except for PWHCs. The most common eco-friendly biological technique used for the enhancement of natural degradation process is bioremediation. Other than biological techniques, some physical and chemical technologies are also developed for the degradation and removal of petroleum contaminants from the environment. Whereas, bioremediation is cheap, eco-friendly, intriguing, and a potential remediation technique for the rehabilitation of oily contaminated soils, and up to some extent it can remediate PWHCs. However, bioremediation is also hampered by several physicochemical, biological, and environmental variables (limiting agents). Therefore, it is difficult to predict the same results of bioremediation in different lab and field conditions. Moreover, little information is available about the genetic, enzymatic, and chemical processes of PWHC’s breakdown in anaerobic conditions. This study contributes to the identification of different effective biological techniques to reduce the harmful impacts of petroleum waste on the environment and health. The present study also describes the role of microbial consortium in combination with nutrients to degrade PWHCs giving ambiguous results, which need further investigation. The knowledge assembled in this review is expected to create new doors for researchers to develop more efficient techniques and control the harmful impacts of PWHCs on the environment and health.

Declarations

Author contribution statement

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The authors declare no conflict of interest.

Additional information

No additional information is available for this paper.

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