A global-scale expert assessment of drivers and risks associated with pollinator decline

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Pollinator decline has attracted global attention and substantial efforts are underway to respond through national pollinator strategies and action plans. These policy responses require clarity on what is driving pollinator decline and what risks it generates for society in different parts of the world. Using a formal expert elicitation process, we evaluated the relative regional and global importance of eight drivers of pollinator decline and ten consequent risks to human well-being. Our results indicate that global policy responses should focus on reducing pressure from changes in land cover and configuration, land management and pesticides, as these were considered very important drivers in most regions. We quantify how the importance of drivers and risks from pollinator decline, differ among regions. For example, losing access to managed pollinators was considered a serious risk only for people in North America, whereas yield instability in pollinator-dependent crops was classed as a serious or high risk in four regions but only a moderate risk in Europe and North America. Overall, perceived risks were substantially higher in the Global South. Despite extensive research on pollinator decline, our analysis reveals considerable scientific uncertainty about what this means for human society.

Animal pollination is key to the reproductive success of >75% of flowering plants globally, including many culturally and economically important plants1,2. Pollination services are estimated to add billions of dollars to global crop productivity and contribute to nutritional security3. Despite these multiple values, there is growing evidence of wild pollinator population declines4,5 and deficits in crop production due to insufficient pollination6, while global demand for pollination services is at an all-time high4 and likely to continue to grow7. Conversely, populations of managed honey bees, while declining in North America and parts of Europe, are increasing in many countries8. Observed trends in wild pollinators have been mostly linked with changes in land management9, climate change10 and agrochemical use11, although these analyses are largely restricted to Europe and North America. Restoring or diversifying habitats and reducing management pressures such as pesticides and grazing have been shown to positively affect wild pollinator populations and managed honey bee health12–15. In response to evidence of declines, pollinators and pollination have attracted public and policy attention globally16–19 and substantial efforts are underway to respond, through national pollinator strategies and action plans18. The Intergovernmental Science-Policy Platform on Biodiversity and Ecosystem Services (IPBES) performed a global assessment of pollinators, pollination and food production from 2014 to 201617. This underpinned the adoption of new commitments to support pollinator conservation by signatories to the Convention on Biological Diversity10 and subsequent steps towards developing national pollinator strategies and action plans in many nations10. One clear message from the pollination assessment was that evidence on the status and trends in pollinator populations, threats and the impacts of their decline, is concentrated in high-income countries, rather than regions thought to be most vulnerable to decreases in pollinator diversity19 and pollination services19. However, unlike the more recent IPBES global assessment on biodiversity and ecosystem services21, the pollination...
assessments did not directly compare and rank the relative importance of major drivers of pollinator decline or make any integrated assessment of the risks it generates for society, either at global or at regional levels. Consequently, although researchers have made broad, global recommendations about how to respond to pollinator decline14, addressing specific drivers and risks at national or regional scales appropriate for policy implementation has been more challenging22.

Here, we used a structured expert elicitation technique and a globally representative group of 20 pollinator and pollination experts, all authors of this paper, to evaluate the relative importance of eight major direct drivers (or causes) of observed pollinator decline and the risks to human well-being associated with ten direct impacts of pollinator decline defined by the IPBES report1 (Tables 1 and 2 and Supplementary Table 1). We separately assessed direct impacts of pollinator decline defined by the IPBES report1, including original wording shown in inverted commas, with section numbers indicated in brackets (Tables 1 and 2 and Supplementary Table 1). We separately assessed direct impacts of pollinator decline defined by the IPBES report1, including original wording shown in inverted commas, with section numbers indicated in brackets (Tables 1 and 2 and Supplementary Table 1). 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We used a semiquantitative risk matrix, with risk scores calculated as the product of probability, scale and severity of impacts and a ‘four-box model’ established by the IPBES (Fig. 1 and Table 3) to communicate levels of confidence1, thus highlighting the key known ‘unknowns’ in current scientific understanding. Our assessment used a modified Delphi technique24, an approach designed to reduce bias but particularly suitable for elicitation of expert judgements about complex issues, where the judgement requires a range of different perspectives and areas of expertise not necessarily held by each participant24.

**Results**

**What is driving pollinator declines?** Figure 2 shows final scores for the importance of the six drivers defined in Table 1, following three rounds of scoring. Globally, land cover and configuration and land management were the most important drivers of pollinator declines (Fig. 2 and Supplementary Tables 2 and 4). Land cover and configuration was scored ‘very important’ in all six regions, while land management was the only variable considered to be ‘the most important’ in any region (Europe) and was ‘very important’ in all
other regions except Africa (Fig. 2). These conclusions are supported by considerable evidence from multiple regions and continuing global trends towards agricultural expansion, conventional intensification and urbanization in regions of the Global South, driven by international trade. Land management was considered less important in Africa, where access to the necessary financial and technical capital to intensify production is still limited and where there was considerable uncertainty (categorized as ‘inconclusive’) over the influence of land cover and configuration (Fig. 2).

Pesticides were scored as ‘important’ or ‘very important’ drivers of pollinator decline in all regions, with the greatest confidence in Latin America and Asia Pacific (Fig. 2). Pesticides were considered less important than land management in Europe and Australia/New Zealand but much more important in Africa (Fig. 2). The adverse effects of pesticides on pollinators have received considerable attention in recent years, following studies demonstrating widespread exposure and detrimental effects on populations or diversity. There is far less evidence available to quantify the exposure in regions beyond Europe and North America. Also, despite very rapid increases in pesticide use since 1990 in middle-income countries of Africa, Latin America and Asia Pacific, pesticide regulations are weaker in the Global South, adding considerably to the risk.

Climate change was considered an ‘important’ or ‘very important’ driver in every region. There was, however, unanimous lack of confidence over its importance relative to other drivers. In every region, except Africa, median confidence scores were ‘medium’ and, in Africa, seven of the ten scorers responded that climate change effects are ‘unknown’ (Extended Data Fig. 2 and Supplementary Table 2). Long-term data scarcity limit and confound the demonstration of current climate change effects on pollinators and available studies are restricted to few taxa such as bumblebees and butterflies.

Genetically modified organisms (GMOs) were considered the least important driver overall, except in Latin America (Fig. 2), which is the second largest producer of GM crops among our regions, after North America. Emerging evidence of potential impacts of herbicide-tolerant crops and associated glyphosate use on honey bees was discussed in the Latin American context (see Supplementary Table 10, now reviewed). Levels of confidence and agreement tended to be low for GMOs and invasive alien species as drivers of pollinator decline, due to very limited available evidence. In the case of GMOs, impacts on pollinators vary according to the type of GM crop and are difficult to separate from the effects of land cover and configuration because such crops are often produced in large monocultures.

**What are the risks to human well-being?** Figure 3 shows the final risk scores following three rounds of scoring, partitioned into probability and magnitude (scale × severity), for each of the direct impacts listed in Table 2, in each major global region. Overall, loss of wild pollinator diversity and crop pollination deficits were

| Table 2 | Direct impacts of pollinator decline on human well-being defined by IPBES, including original wording from Table 6.2.1 of ref. 1 shown in inverted commas |
|---|---|---|
| Impact | Definition | Example |
| **Impacts on food production** | | |
| Pollination deficits | ‘Crop pollination deficit leading to lower quantity or visual/nutritional quality of food (and other products…)’. | Reduction in the quantity or quality of food, fibre, fuel or seed that can be produced, as a result of pollinator loss. |
| Yield instability | ‘Crop yield instability due to loss of pollinators or change in pollinator communities’. | Crop yields becoming less stable or predictable between years or locations. |
| Honey production | ‘Fall in honey production (and other hive products)’. | Reduction in the amount of honey or hive products that can be produced, as a result of pollinator loss. |
| Food system resilience | ‘Decline in long-term resilience of food production systems’. | Resilience is the ability of the food production system to withstand or recover from shocks or adverse effects, such as changes in climate. |
| Wild fruit availability | ‘Decline in yields of wild fruit, harvested from natural habitats by local communities’. | Fruits or seeds harvested for food by people (not by animals). Could include, for example, blueberry harvesting from wetlands or Rubus fruticosus fruits harvested from hedgerows. |
| Managed pollinators | ‘Reduced availability of managed pollinators’. | Managed pollinators are animals used to provide crop pollination, rather than for the production of honey. |
| **Impacts on biocultural diversity** | | |
| Wild pollinator diversity | ‘Loss of wild pollinator diversity’ leading to long-term changes in network/food web interactions. | Loss of species richness or abundance of particular species of wild pollinators, including invertebrates and vertebrates. This impact is intermediate; ultimate impacts on human well-being can include food system resilience, aesthetic value, cultural practices and traditions. |
| Wild plant diversity | ‘Loss of wild plant diversity due to pollination deficit’. | Loss of species richness or abundance of particular species of wild plants due to pollination deficit. This impact is intermediate; ultimate impacts on human well-being can include loss of ecosystem services such as erosion prevention, aesthetic value, cultural practices and traditions. |
| Aesthetic values | ‘Loss of aesthetic value, happiness or well-being associated with wild pollinators or wild plants dependent on pollinators’. | This could include amenity values of specific plant communities, values of emblems or symbols and the value of pollinators as sources of inspiration for art, music, literature, religion and technology. |
| Cultural values | ‘Loss of distinctive ways of life, cultural practices and traditions in which pollinators or their products play an integral part’. | Cultures, traditions and behaviours involving pollinators or pollinator products. This includes beekeeping, honey-hunting and specific dances or rituals associated with pollinators. |

For a definition of ‘biocultural diversity’ in this context, see Hill et al.26.
the highest and most widespread risks, scoring as serious or high risks in every region (Fig. 3 and Supplementary Tables 3 and 7). Although much of the published evidence for pollinator declines is from Europe and North America (where the evidence was considered ‘well established’); there is growing evidence of pollinator declines in other regions, including vertebrate pollinators, along with global evidence of general biodiversity decline. Evidence for crop pollination deficits is also growing across several regions, and Latin America is also home to a high diversity of extant indigenous cultures and people, many of whom rely on subsistence agriculture and natural resources such as non-timber forest products, increasing the risks from a decline in honey, wild fruits and cultural values.

In contrast to Latin America, Africa had very low risk scores for honey production and managed pollinators (both ‘low’ risk; Fig. 3 and Supplementary Table 3). Beekeeping is unique in Africa since it is the only global region that has large, genetically diverse populations of native honey bees (Apis mellifera, various subspecies) still thriving in the wild. In fact, numbers of managed hives are increasing in many African countries due to limited colony losses and managed honey bee populations relatively resilient to Varroa mite. The risk of loss of aesthetic values, happiness or well-being associated with wild pollinators or wild plants dependent on pollinators was perhaps the most difficult to score in all regions. In some contexts, one can make an argument that aesthetic values associated with pollinators are increasing, as people become more aware of their roles, beauty and diversity. Discussions focused on what constitutes aesthetic values and how they might be changing in response to pollinator decline (Supplementary Table 11). This risk varied regionally, with Latin America and Africa scored highest (42) and lowest (4) risk, respectively (Fig. 3 and Supplementary Table 3). While clear links exist between people and pollinators or pollinator-dependent plants in both regions, for Latin America, these links are often specific to the large-scale areas, such as the flower-rich shrubland of Namaqualand, southern Africa, making potential impacts of pollinator decline on aesthetic values less clear (Supplementary Table 11).

Europe was the region where human well-being was considered at the lowest risk from pollinator declines overall (mean risk score = 19.6), with no ‘high’ risks and only two ‘serious’ risks (pollination deficits and wild pollinator diversity). Unlike Latin America, many European countries grow relatively few crops that are highly pollinator dependent and food systems, particularly within the European Union, are highly industrialized and globalized, greatly reducing the importance of wild fruits and buffering against the impacts of global change on food system resilience (both ‘low’ risk). Despite evidence that habitats containing pollinator-dependent plants are aesthetically valued in Europe, their cultural importance may be lower than elsewhere in the world, although this was

| Table 3 | Communication of the degree of confidence |
|---------|------------------------------------------|
| Confidence category | Definition | Thresholds, based on third round modified Delphi scores |
| Well established | Robust evidence | Confidence score ≥66.7% and proportion unknowns <40% |
| High agreement | For risks, IQR ≤ 3; for drivers, IQR ≤ 1 |
| Established but incomplete | Low-quality evidence | Confidence score <66.7% or ≥40% of responses ‘unknown’ |
| High agreement | For risks, IQR ≤ 3; for drivers, IQR ≤ 1 |
| Unresolved | Robust evidence | Confidence score ≥66.7% and proportion unknowns <40% |
| Low agreement | For risks, IQR > 3; for drivers, IQR > 1 |
| Inconclusive | Low-quality evidence | Confidence score <66.7% or ≥40% of responses ‘unknown’ |
| Low agreement | For risks, IQR > 3; for drivers, IQR > 1 |

We follow the four-box model for the qualitative communication of confidence (Fig. 1). The degree of confidence in each finding is based on the quantity and quality of evidence, represented by confidence scores (Methods) and level of agreement among scorers, represented by interquartile ranges (IQRs) of expert scores for each variable.
highly uncertain, with our risk score for ‘cultural values’ in Europe categorized as ‘inconclusive’ due to low confidence and low agreement among scorers.

Loss of access to managed pollinators was only considered a serious risk to people in North America, where honey bees *A. mellifera* represent a key input to large-scale, industrialized cropping systems such as almond51 and have suffered serious declines in the past due to outbreaks of disease, pests and ‘colony collapse disorder’52. The probability of the same occurring in, say, Latin America or Asia Pacific, was considered far lower, even if the severity of the impact would be similar (Fig. 3 and Supplementary Table 3). Experts were divided (low agreement) on the risk from losing managed pollinators in Europe (Fig. 3), where markets for pollination services are less well developed than in North America53, and in Latin America, where the number of managed honey bee colonies has expanded substantially but pressures on their populations remain high7.

Across both risks and drivers, there was high agreement but low confidence for most factors, placing them in the ‘established but incomplete’ confidence category. Our confidence in several direct impacts was low because of numerous gaps in knowledge about the ecology and status of all but the most common pollinator species and the relationships between pollinators, human economies and culture52,54. Furthermore, while statistical information on crop production, managed pollinators and honey production is often collected at a national scale, the quality of these data varies considerably within a region and over time and does not capture subsistence agriculture, particularly in the Global South.

**Discussion**

In our analysis, the global ranking of drivers of pollinator decline by importance (Fig. 2) differs from the order of relative impact of direct causes of biodiversity loss (or ‘changes in the fabric of life’) presented by Diaz et al. on the basis of the IPBES global assessment41. In both cases, land use change (here, land cover and configuration) for terrestrial realms is the most important driver but, for the whole of nature42, ‘direct exploitation’ is the next most important driver, followed by climate change, pollution and invasive alien species. For pollinators, direct exploitation is broadly equivalent to ‘pollinator management’ (not including direct harvesting of pollinators or pollinator products, which is not suggested as a major driver of pollinator decline). This was ranked with lower importance than climate change, pesticides and pests and pathogens in our assessment. For pollinators, climate change was ranked below pesticides as a driver, perhaps reflecting more complete evidence that current
pollinators are likely to be longer term. Much of the current evidence shows shifting ranges, which only sometimes translate into population declines or highly uncertain projected future distributions under climate change. Although these two analyses used different methods for ranking drivers (Diaz et al. quantified the relative impact of each driver on the basis of rankings in published studies comparing two or more drivers), it is not surprising that the relative importance of drivers differs, when focusing on a functionally defined subset of organisms (pollinators) that are almost all relatively small in size.

Despite high-profile, extensive research on the drivers and impacts of pollinator decline, our analysis reveals considerable scientific uncertainty about what this means for human society, regionally and globally. There are clear risks of wild pollinator diversity loss and crop pollination deficits globally, yet less is understood about the broader implications for human well-being. The case for action to address pollinator decline is most clearly made for Latin America (Fig. 3).

We followed an explicit, transparent and systematic process of risk assessment, as recommended by Zommers et al. for robust climate change risk assessment. Even so, a number of limitations to this approach have been clearly defined. Perhaps the most pertinent here is the potential for our results to be influenced by the value judgements and world views of our individual experts. For example, when rating ‘severity’ of impacts, people whose lives are directly affected might be inclined to rate severity more highly than those unaffected. When rating ‘probability’, interpretation of verbal scales by individual experts can be poorly aligned or even overlap, when measured against numerical scales; in extreme cases, what is ‘likely’ to one person can be considered ‘unlikely’ by another. One way to reduce this subjectivity would be to define explicit, sharp or fuzzy boundaries for the categories in our verbally described scales (Supplementary Table 1), by using numerical scales, thereby reducing ‘vagueness’. Several underlying numerical scales can be conceptualized for all the elements of risk we assessed. Possible scales could be derived from available data on the impacts themselves over time or space (for probability), the numbers or proportions of people who could be affected (for scale) and contributions to health, well-being and income from particular activities (for severity).

For example, to judge the probability of a fall in honey production, we discussed the relevance and quality of available data on honey production and numbers of managed honey bee hives, and the trends shown by these datasets, for each region. To judge the scale of impact of a fall in honey production in terms of numbers of people affected, we considered numbers of beekeepers, honey hunters and honey consumers, across each region. To judge the severity of this impact, we considered the proportions of beekeepers’, farmers’ and honey hunters’ incomes that come from honey, and the relative impacts of honey on people’s individual health outcomes (Supplementary Table 1). However, for most of our impacts, numerical data were available only for a small proportion of the issues considered, in a subset of possible contexts and usually not at regional scale, so using numerically specified boundaries would still have demanded subjective judgements or speculation. In these circumstances, providing numerical scales to delineate the categories would represent an unfounded and misleading level of precision.
Our process reveals several major knowledge gaps. There is an urgent need for research in Africa\(^1\), to address the substantial uncertainties around the risks to people from pollination deficits (Fig. 3), and the importance of changes in land cover and configuration, as a driver of pollinator decline (Fig. 2). In more developed regions, especially North America, we lack understanding of the scale and severity of impacts of pollinator decline on human well-being (Supplementary Table 3). Globally, the consequences of climate change for pollinators and pollination remain poorly understood but its impacts will clearly increase in prominence in the coming decades\(^2\). As climate change is very likely to interact with other drivers of pollinator decline, a focus on how to mitigate and adapt to it should be central to pollinator research and conservation strategies.

**Method**

We assessed drivers and risks using a modified version of a formal consensus method known as the Delphi technique\(^3\), in which the second and third rounds of anonymous, independent scoring took place following detailed discussions at a face-to-face workshop in November 2017. This modification of the Delphi technique is frequently used in environmental research, where issues are multidisciplinary and interpretations of the same phrase can differ strongly among individuals\(^4\). All but one of the authors of this paper (referred to here as ‘expert’) took part in all rounds of the Delphi process (D.S. facilitated only and did not score). This set of 20 pollination experts was carefully selected to cover the range of necessary expertise, including biodiversity science, economics, social science and indigenous and local knowledge, and to ensure that the main global regions were each represented by at least two scorers either originating from or mainly working in that region. Thirteen of the 21 authors (62%) were also authors of the IPBES global pollination assessment\(^5\), mostly nominated by their respective national governments, and the team had a balanced gender ratio of 11 men to ten women.

**Definitions of regions, parameters and scores.** We divided the world into six global regions, largely representing geographic continents of North America, South America, Asia, Europe, Africa and Oceania, with two key differences: (1) we included the Pacific islands in a region known as ‘Asia Pacific’, rather than combining them with Australia and New Zealand in the geographic continent ‘Oceania’. Our Asia Pacific region is equivalent to most of the Asia Pacific as defined by IPBES but excludes Australia and New Zealand. We named ‘Australia/New Zealand’ as a separate region because they are very different from mainland Asia and the Pacific islands both biogeographically and geopolitically (Supplementary Fig. 1); (2) we included the countries of Central America and the Caribbean with Latin America, rather than with North America as they would be in the geographic continent. Our ‘Latin America’ region includes the subregions of Mesoamerica, the Caribbean and South America, as defined by IPBES (Supplementary Fig. 1).

For each region, experts individually assigned probability, scale and severity scores for each of ten impacts of pollinator decline and importance scores to each of eight drivers of pollinator declines defined by the IPBES (Tables 2 and 3, respectively), using the five-point Likert scale described in Supplementary Table 1. All scores were accompanied by a confidence score of low, medium or high, enabling experts to qualify their judgements with a level of confidence, on the basis of the amount of evidence they were aware of and its quality.

The following definitions of probability, scale and severity were available for authors to consult throughout the process:

- **Probability:** A high probability of impact suggests that the impact is already taking place or is very likely, at least in some circumstances. Low probability implies that the impact is not taking place or is unlikely. Unknown means there is not enough evidence to make a judgement on whether or not the impact is happening or likely to happen.
- **Scale of impact** refers to the numbers of people or to the area affected. Large means there is evidence for impacts on people and livelihoods, either over a large area or affecting many people. Moderate means there is evidence for impacts on people and livelihoods, either over a moderate area or affecting a moderate proportion of people. Small means there is evidence for impacts on people and livelihoods, either in a small, localized area or only affecting a small number of people. Unknown means there is not enough evidence on the scale of this impact to make a judgement.
- **Severity of impact** refers to the nature of the impact on individual people or families. Large means there is evidence for a substantial or severe impact on people and livelihoods. Moderate means there is evidence for a moderate impact on people and livelihoods and small means a small impact. Unknown means there is not enough evidence on the severity of this impact to make a judgement.

Experts rated the importance of each driver in affecting pollinators, at the present time, in each specific region, on a 1–5 scale from ‘not important’ to ‘the most important’ (Table 1 and Supplementary Table 1).

We set an a priori expectation of consensus as an interquartile range of <2 between scores for a particular element (not including confidence). This still allowed us to distinguish between high and low agreement following the outcomes described in Table 3, in which high agreement was defined as an IQR ≤ 1 (where half of all scores are the same or an adjacent score) (Table 3).

**Three iterative rounds of scoring.** In an initial scoring phase, all experts were invited to comment on the proposed scoring structure described above. Following this, the first round of scoring was conducted online in October 2017. Each expert was asked to score for all regions, considering the evidence in the IPBES report alongside their own expertise. Experts could add comments to support their scores and were encouraged to cite parts of the IPBES report and other specific literature. Scores and comments were compiled anonymously and summaries sent to all experts, detailing the median and interquartile range of scores for each element and the proportions of ‘unknown’ responses.

In November 2017, all experts attended a 3-day, face-to-face workshop in Reading, UK. Experts were divided into two groups, which each discussed the results from the first round, and the evidence that supports them, for three regions. Group 1 discussed and scored in rounds 2 and 3 for Europe, North America and Africa; group 2 discussed and scored Latin America, Asia Pacific and Australia/New Zealand. Discussions were facilitated around a single theme. Facilitators kept in contact and discussed any specific issues arising about how to score, to ensure that both groups responded in the same way. At the end of each part of the discussion, participants scored again for each element of risk and each driver, for each region in turn. Scoring was conducted independently and anonymously, using Excel spreadsheets on personal laptops. All members of a group were encouraged to score for each region discussed in their group, with the following guidance: ‘Score if you can (but you don’t have to). If you feel confident to score for a region outside your own personal knowledge, please do so. These issues are complex and open to interpretation. This is why we employ a subjective scoring process, with anonymous scoring. Listen to the discussion, and then score as you understand it’.

These round 2 results were compiled as before and any scores with IQR ≥ 2 (our a priori criterion for consensus), progressed to round 3 for resoring.

Round 3 scoring took place on the third day of the workshop in a plenary discussion. This allowed a further opportunity for any consistent differences in scoring or approach across groups to be revealed but none was evident. Second round scores were presented and made the subject of debate and discussion. Experts scored again anonymously and independently, using laptops, for the regions they scored for in round 2, although the discussion was open to both groups. In total, 19 variables (three drivers and 16 impacts) were rescored, along with associated confidence levels. Due to an error affecting two Latin American regions (Latin America; pollination deficit (severity), yield instability (scale), wild fruit availability (scale) and wild plant diversity (scale)) with IQR ≥ 2 were not flagged for resoring during the workshop and were later resored during a teleconference. Only five of the ten scorers from group 2 were able to attend the teleconference, due to time differences, so these four variables have only n = 5 scorers (the final dataset (Supplementary Fig. 3)). All other variables have at least eight scorers. Following the third round, three variables still failed to reach consensus (IQRs ≥ 2) — Australia/New Zealand: pollination deficit (probability), wild fruit availability (probability) and Latin America: managed pollinators (probability) (Supplementary Table 3).

**Analysis.** Median scores following the third round of scoring were used to derive risk scores (the product of probability, scale and severity scores) and associated risk categories (boundaries visualized in Fig. 3), importance scores for drivers and confidence categories for all final scores, following criteria given in Table 3. In assigning confidence categories, the quantity and quality of evidence was based on assigned confidence scores for each risk or driver. The confidence score is the percentage of the maximum possible confidence score (9 for risks, 3 for drivers) represented by the median confidence scores from the final round, with the three medians summed in the case of impacts (confidence score for risk = (Σ confidence scores for probability, scale and severity)×100)).

Overall global scores for the importance of drivers were calculated as a median of the region-level scores and comitance scores, to ensure equal weight was given to each region (although the numbers were unchanged if individual scores across all six regions were used). We did not calculate overall global risk scores for different impacts of pollinator decline because these scores were based on assessments of probability, scale and severity for different global regions and it does not make sense to average these across regions. All figures were drawn using the ‘ggplot2’ package\(^6\) in R v4.0.0.

We proposed that the scores participants gave for each component of the risk or driver importance were dependent on the impact or driver being scored and on
the region being scored, rather than reflecting individual scorer differences. We tested this hypothesis using cumulative link models and cumulative link mixed models with logit link functions (also called proportional odds or ordinal logistic regression models), with the ordinal package in R v.4.0.3. The top and bottom two score categories (scores 1 and 2 and 4 and 5, respectively) were collapsed to create three-point scales for probability, scale, severity and importance of drivers.

We considered the effect of region and impact, or region and driver, on score for each of four dependent variables: probability, scale, severity and importance. ‘Unkown’ responses were treated as missing values for this analysis. The dataset was not large enough to examine the interaction between region and impact or driver with this type of model (n ≤ 10 scorers for each combination of factors).

For each model, we tested the proportional odds assumption that the effects of region or impact group were the same, regardless of where the cut-off points were placed across the three score categories, using the nominal test and scale test functions, which use likelihood ratio tests. When this assumption was violated, we used partial proportion odds models where possible, given our data structure. Independent variables that failed the tests were examined, with scale (dispersion of latent variable) allowed to vary among levels of the dependent variable (failure of the scale test) or effects of the relevant factor assumed to be nominal rather than ordinal (failure of the nominal test).

These models do not account for the random effects of scorer or group because the scorers were divided among two separate groups, each of which only scored half of the regions. We ran cumulative link mixed models separately for each group, including scorer as a random effect to account for differences between individual scorers. The effects of group cannot be analysed as a random factor with this study design because there are only two levels. The effect of group cannot be separated from the effect of region in a single model.

We used McFadden’s pseudo $R^2$ value ($\rho^2$) to provide an indication of goodness-of-fit for all models, as recommended by Menard . This is calculated relative to a null model using the following equation:

$$\rho^2 = 1 - \frac{LL_{mod}}{LL_0}$$

where $LL_{mod}$ is the log likelihood value for the fitted model and $LL_0$ is the log likelihood for the null model which includes only an intercept as predictor (so that every score is predicted with the same probability).

Results of this analysis are provided and discussed in the Supplementary Information (Supplementary Tables 4–9 and accompanying text).

Reporting Summary. Further information on research design is available in the Nature Research Reporting Summary linked to this article.

Data availability

Figures 2 and 3 represent scores from round 3 of a Delphi process with n = 20 expert scorers. Medians and interquartile ranges for these scores are presented in full in Supplementary Tables 2 and 3; the raw data are shown in Extended Data Figs. 2 and 3.

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**Author contributions**

L.V.D. conceived and designed the study. L.V.D and T.D.B. contributed equally to data collection, analysis and writing the paper. S.G.P and H.T.N.T. convened the expert panel. S.G.P., D.S., T.D.B., H.T.N. and L.V.D. designed, organized and ran the workshop. L.V.D., T.D.B., H.T.N., A.J., M.A.A., P.B., D.B., L.G., L.A.G., B. Gemmill-Herren, B. G. Howlett, V.L.I.-F., S.D.J., A.K.-H., Y.J.K., H.M.G.L., T.L., C.L.S., A.J.V. and S.G.P. contributed to all rounds of scoring and discussion and commented on edited the final manuscript. D.S. contributed to discussions and commented on edited the final manuscript.

**Competing interests**

The authors declare no competing interests.

**Additional information**

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Extended Data Fig. 1 | Definition of global regions according to biogeographical and geopolitical conditions. Definition of global regions according to biogeographical and geopolitical conditions.
Extended Data Fig. 2 | Final breakdown of scoring of direct drivers by world regions and importance. Final breakdown of scoring of direct drivers by world regions and importance.
Extended Data Fig. 3 | Final breakdown of scoring of risks by world regions, impact and components of risk (probability, scale, severity). Final breakdown of scoring of risks by world regions, impact and components of risk (probability, scale, severity).
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| Study description | This study is an expert elicitation, following a modified Delphi process. 20 purposively selected participants discussed evidence and anonymously submitted scores reflecting their judgement on 10 impacts of pollinator decline on human well-being, and eight drivers of pollinator decline, in a three-round iterative process. |
| Research sample | Participants were 20 purposively selected pollinator and pollination experts, all authors of the paper (therefore not research subjects, but collaborators). Thirteen of the 20 authors were also authors of the IPBES global pollination assessment, mostly nominated by their respective national governments, and the team had a relatively balanced gender ratio of 11 men : 9 women. |
| Sampling strategy | The set of 20 pollination experts was purposively selected to cover the range of necessary expertise, including biodiversity science, economics, social science and indigenous and local knowledge, and to ensure that the main global regions were each represented by at least two scorers either originating from or mainly working in that region. |
| Data collection | Data were recorded by individual participants in an Excel spreadsheet on their own laptops, and transferred to a master spreadsheet by author Tom Breeze, during the workshop. Quality control involved cross-checking individual spreadsheets against the master dataset, in randomly selected data points and summary data, in both R and Excel. |
| Timing | Data were collected from October to November 2017, in the year following publication of the IPBES pollinators and pollination assessment. |
| Data exclusions | No data were excluded from analysis. |
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