Co-Pyrolysis of Waste Printed Circuit Boards With Iron Compounds for Br-Fixing and Material Recovery

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Co-pyrolysis of waste printed circuit boards with iron compounds for Br-fixing and material recovery

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Abstract Waste printed circuit board was co-pyrolyzed with iron oxides and iron salts. Solid, liquid and gaseous products were collected and characterized. Co-pyrolysis with FeCl\textsubscript{2}, FeCl\textsubscript{3} or FeSO\textsubscript{4} was able to increase the yield of liquid product which was rich in phenol and its homologues. Also, the addition of co-pyrolysis reagents reduced the release of brominated organics to liquid as Br was either fixed as FeBr\textsubscript{3} in solids or released as HBr. In particular, FeCl\textsubscript{2} showed the best ability to reduce the release of Br-containing organics to liquid compared with FeCl\textsubscript{3} and FeSO\textsubscript{4}. Solid residuals were rich in iron oxides, glass fibers and charred organics with surface areas of 20.6-26.5 m\textsuperscript{2}/g. CO\textsubscript{2} together with a small amount of CH\textsubscript{4} and H\textsubscript{2} were detected in the gaseous products. Overall, co-pyrolysis could improve the quantity and quality of liquid oil which could be reused as chemical or energy sources. Pyrolysis of waste printed circuit board was promising as a method for recycling.

Keywords: waste printed circuit board, co-pyrolysis, iron compounds, Br-fixing, liquid yield

1. Introduction

Generation of electronic waste worldwide has been growing consistently in recent years due to rapid technological advancements and short life span of electrical and electronic equipment (Zhu et al. 2020). As an import part of electrical and electronic equipment, printed circuit boards contained resources of high value including copper, tin, gold and silver etc. In general, it contained 20-30% metals, about 20% plastics while ceramics and glass fiber making up the rest (Nekouei et al. 2020).

Traditionally, after electronics reached their life span, the waste printed circuit
boards (WPCB) were separated and undergone a rough recovery of profitable metals (Cu, Au etc.) before they were disposed by landfill or incineration which could lead to serious contamination of air, soil and water (Mdlovu et al. 2018). In recent years, China has tightened its regulations on WPCB disposal after many reports of pollution by toxic metals and hazardous organic substances remained in WPCB. Effective recovery of metal resources (such as iron and copper) and non-metal fractions (such as resins) could reduce not only the environmental impact of WPCB but also the demand on natural ores (Hao et al. 2020).

One of the difficulties in WPCB recycling was its complex composition. It contained a wide range of organic and inorganics which varied according to device type, manufacturers etc. Bazargan (2014) reported that WPCB could contain up to 40 kinds of metals (e.g. Cu, Fe, Al, Sn, Ag, Au) and 10 kinds of non-metals. Some of the metal contents were even higher than mineral ores. Therefore, WPCB had great economic potential. Recovery of metals was usually more attractive due to it high added-value.

With the advancement in research, the less valuable non-metallic fraction could also be recycled as a source of materials and energy (Ghosh et al. 2015; Shen et al 2018).

For recycle, WPCB was first disassembled manually or automatically to remove batteries and capacitors components. Then it was shredded, crush or ground/pulverized before physical, chemical and biological processes to further separate and enrich metals from resin, fiber glass and plastics (Shen et al. 2018). Physical separation methods could be shape-, density-, conductivity- or magnetism-based (Salbidegoitia et al. 2015). Other processes investigated extensively to recycle metals or non-metals included
gasification (Shen et al. 2018), supercritical fluid treatment (Golzary and Abdoli 2020),
dissolution in organic solvents (Yousef et al. 2018), hydrometallurgical and
pyrometallurgical techniques (Zhou et al. 2011; Qiu et al. 2020), pyrosis (Kim et al.
2018) and hydrothermal treatment (Yin et al. 2011).

As a waste disposal method for WPCB, pyrolysis was able to recycle both metals
and organic matters. It was carried out under non-oxygen atmosphere and organics
were decomposed into small molecular products. Metals was not oxidized under inert
atmosphere and became easier to be liberated after pyrolysis while organics were
converted to fuel or other valuable chemicals (Liu et al. 2021). Kim et al. (2015)
reported that phenols and bisphenol A are the dominant compounds in the liquid phase
products in pyrolysis of epoxy-PCB and paper-laminated phenolic printed board.
Rajagopal et al. (2016) used physical activation on char obtained from pyrolysis and
obtained activated carbon with a surface area as high as 700 m²/g. Also, pyrolysis was
carried out under relatively low temperature and normal pressure which means low
energy consumption and cost (Shen et al. 2018).

However, WPCB contained many halogenated compounds as flame retardant.
Among them, the most common is brominated compounds. The degradation of these
compounds was found to produce toxic products such as dibenzo-p-dioxin, dibenzo-
furan, 2-bromophenol, 2,6-dibromophenol and methyl bromide etc. (Ning et al. 2017).
Co-pyrolysis with various additives were investigated to minimize the contents of
brominated compounds in pyrolysis oil and gas (Ma et al. 2018). KOH, NaOH, K₂CO₃,
NaCO₃, CaCO₃, La₂O₃, MgO, CuO, CaO, Al₂O₃, zeolite, iron oxides, carbon powder
have all been used as debrominated agents (Evangelopoulos et al. 2015). Shen (2020) reported that pretreatment of WPCB by alkali was better in Br fixing than alkali-salt. Br-fixing efficiency reached 53.6% by NaOH as Br released was adsorbed by NaOH to form NaBr and was retained in chars. Terakado et al. (2011) compared the debromination effects of metal oxides. The ability of oxides on HBr emission suppression varied. La$_2$O$_3$ and CaO were more effective with a suppression efficiency of about 90%. Br-fixing was achieved by bromination of oxides as LaOBr, CuBr$_2$, CaBr$_2$, FeBr$_2$ were found in the solid residues of pyrolysis. Many studies have shown that by choosing the right reagent, co-pyrolysis could not only reduce Br release into pyrolysis products but also induce higher yield of valuable oil or change the properties of solid residues (Kakria et al. 2020).

This research attempted to study the feasibility of WPCB recycling by co-pyrolysis with iron oxides and iron salts. Studies have shown that iron is a good catalyst for the graphitization of carbonaceous compounds. Wu et al. (2018) reported that Fe$^{2+}$ or Fe$^{3+}$ salts can be used as template and catalyst in their pyrolysis of phenol formaldehyde resin and ethylene glycol. Activation and pyrolysis worked together when FeCl$_2$ and FeCl$_3$ were added, leading to increase in yield and complete carbonization.

The WPCB herein was from Jiangsu Province, China. It was already disassembled, crushed and shredded with most of the electronic components and metals recovered. The residual WPCB, mainly non-metallic, was currently used as an additive in brick-making. However, the amount that could be consumed this way was limited by the demand for bricks. An additional outlet for the WPCB was urgently needed as landfill.
was no longer an option while incineration was problematic in fear of dioxin emission.

This research focused on co-pyrolysis of WPCB with iron oxides and iron salts for the possibility of energy or materials recovery. Conditions of pyrolysis (temperature, mass ratio etc.) on the pyrolysis products were investigated in detail. In particular, the effects of different reagents (Fe$_3$O$_4$, FeCl$_2$, FeCl$_3$ and FeSO$_4$) on liquid yield, and characteristics of residual solids and liquid products were compared. Behavior of WPCB and reagents during pyrolysis were studied via the monitoring of crystalline phases, morphology, surface chemistry and elemental compositions. In addition, compositions of liquid products were investigated to study the Br-fixing behavior of different reagent.

2. Materials and methods

2.1 Materials

Waste printed circuit board (WPCB) used in this study was from a printed circuit board recycling company in Kunshan City, Jiangsu Province, China. Electronic components were disassembled first and a rough metal recovery was conducted by crushing and sorting. Therefore, the raw material in this study was mainly non-metallic with a residual metal content of 15.51%. Proximate analysis showed that WPCB contained 29.62% volatile carbon, 4.52% fixed carbon and 65.86% of ash.

Fig. 1 is a SEM image of the WPCB. It consists of particles and cylinder-shaped glass fiber. Particle size of the WPCB were mostly in the range of 10-200 μm.
Other materials used in the study, including FeSO$_4$·7H$_2$O, FeCl$_2$·4H$_2$O, FeCl$_3$·6H$_2$O, and Fe$_3$O$_4$ were all purchased from Sinopharm Chemical Reagent Co., Ltd. (Shanghai, China).

2.2 Pyrolysis

Pyrolysis of WPCB were carried out in a laboratory-scale tube furnace (OTF-1200X, Kejing, China) under nitrogen atmosphere. The reaction tube is 1000 mm in length and 54 mm in diameter. Fe$_3$O$_4$, FeSO$_4$, FeCl$_2$ and FeCl$_3$ were selected as reagents for co-pyrolysis. Co-pyrolysis of WPCB was carried out at temperatures ranging from 350-750°C. Fe$_3$O$_4$ in its powder form was mixed directly with WPCB. For co-pyrolysis with FeCl$_2$, FeCl$_3$ and FeSO$_4$, 10 g of WPCB were added to different volumes of solutions of 5 g/L then freezing dried to remove water. The FeCl$_2$, FeCl$_3$ and FeSO$_4$ to WPCB mass ratios were set at 0.25, 0.50 and 1.0. Before heating, the furnace was purged for air with N$_2$ at a rate of 10.0 mL/min for 30 min. Temperature was raised at a rate of 10°C/min. Time of pyrolysis was kept constant at 60 min. The outlet of tube furnace was connected to a condenser with cooling system to collect liquid oil. The remaining gas passed through dilute NaOH solution before exhaust was collected. All pyrolysis
samples were done in triplets. Results given were the averages.

Liquid yield (in percentage) was calculated as the mass of liquid collected divided by mass of initial WPCB. Co-pyrolyzed sample was named as WPCB/reagent-temperature, such as WPCB/FeCl$_2$-650.

2.3 Characterization of pyrolysis products

Thermogravimetric analysis of WPCB was conducted via a Thermal Gravimetric Analyzer (TGA, STA-8000, PerkinElmer, USA). X-ray diffraction analysis was carried out by X-ray Diffractometer (Rigaku Ultima IV, Japan) with a Cu-K$\alpha$ radiation at 40 kV and 30 mA as 2$\theta$ ranged from 10° to 80°. FTIR (NICOLET iS10, ThermoScientific, Germany) analysis was employed to characterize functional groups of solids. Morphologies of solid particles were analyzed by Scanning Electron Microscopes (Nova Nano 450, FEI and ZEISS Gemini 300, Germany). Proximate analysis of WPCB was carried out according to ASTM standard method E1131-08. Composition of pyrolysis liquid was analyzed via a Gas Chromatography-Mass Spectrometry (GC-MS, 7890/5975C, Agilent, USA). GC-MS was operated with a HP-5MS capillary column (30m×0.32 mm×0.25m). The initial temperature of the oven is 40°C and kept for 5 min before temperature was raised to 260°C at 5°C/min and keep at 260°C for 20 min. The carrier gas was helium with a constant flow of 1.0 mL/min.

3. Results and discussion

3.1 Screening of co-pyrolysis reagent by liquid yield

In this research, WPCB and WPCB mixed with iron compounds were pyrolyzed at high temperature to investigate the effects of these additives. The compounds selected
were Fe$_3$O$_4$, FeCl$_2$, FeCl$_3$ and FeSO$_4$.

Fig.2 is the change in liquid yield of pyrolysis with the change of temperature and iron compound/WPCB ratio.

Studies have shown that liquid products from WPCB pyrolysis could be a valuable source of chemicals or energy as it contained products such as phenols and bisphenol A etc. (Quan et al. 2010; Gao et al. 2020). Fig.2(a) shows how liquid yield varied with temperature when different iron compounds were used. Iron compound/WPCB ratio was kept constant at 0.5. Liquid yield peaked at 550$^\circ$C for WPCB. The addition of Fe$_3$O$_4$ actually led to a slight drop in yield per unit of WPCB. In contrast, the addition of FeCl$_2$, FeCl$_3$ and FeSO$_4$ was beneficial probably due to these iron salts' ability to catalyze graphitization of carbonaceous materials. The highest yield increased from 13% to around 18-20%. The optimal pyrolysis temperature was 550-650$^\circ$C.

Iron compound/WPCB ratio also affected liquid generation (Fig. 2(b)). The pyrolysis temperature was set at 650$^\circ$C for FeCl$_2$ and 550$^\circ$C for the rest. There is a
general trend of increase in yield with the increase of mass ratio from 0.25 to 0.5. Higher ratio means more iron mass per unit of WPCB. The effects of mass levelled off after the ratio reached beyond 0.5 for FeCl₂, FeCl₃ and FeSO₄. This is in accordance with results from Gao et al. (2019). They also found there is an optimal mass ratio between iron oxide and WPCB for liquid yield from pyrolysis.

Next, pyrolysis behavior of WPCB with FeCl₂, FeCl₃ and FeSO₄ was investigated in detail.

3.2 Pyrolysis with FeCl₂

Fig. 3 is the mass loss of WPCB, FeCl₂ and WPCB/FeCl₂ during pyrolysis at different temperature and XRD, FTIR and SEM images of the residual solids. The FeCl₂/WPCB ratio was set at 0.5. Fig. 3(a) is the mass loss during pyrolysis. 20.8% of mass was lost when WPCB was pyrolyzed at 350°C for 1 h. As pyrolysis temperature was raised from 450°C to 750°C, mass loss increased only slightly from 22.6% to 24.7%. For comparison, mass loss of FeCl₂ in N₂ was also recorded in Fig. 3. As temperature was raised, FeCl₂ was first converted to FeOOH which can further decomposed into iron oxides such as Fe₂O₃ with reactions (1) and (2) (Kang et al. 2010).

\[
\begin{align*}
120 – 200°C: & \text{FeCl}_2 \cdot 2\text{H}_2\text{O} \rightarrow \text{FeOCl} \cdot \text{H}_2\text{O} + \text{HCl} \rightarrow \text{FeOOH} + 3\text{HCl} \\
200 – 490°C: & 2\text{FeOOH} \rightarrow \text{Fe}_2\text{O}_3 + \text{H}_2\text{O}
\end{align*}
\]
Fig. 3 Characterization of WPCB, FeCl$_2$ and WPCB/FeCl$_2$: (a) mass loss; (b) XRD; (c) FTIR; (d) SEM image of WPCB/FeCl$_2$-650

So mass loss was caused by FeCl$_2$ being converted to iron oxides. There is a sharp mass loss of FeCl$_2$ between 350-550°C. Calculated mass losses in Fig. 3 were obtained assuming mass change of WPCB and FeCl$_2$ in WPCB/FeCl$_2$ was independent of each other. Mass losses labelled as WPCB/FeCl$_2$ were the real values of WPCB/FeCl$_2$ mixture. Total mass loss ranged from 10.9-33.6% as compared to the calculated loss of 7.4-29.9%. Real mass losses were slightly higher than calculated ones. The higher mass loss could be due to interactions between WPCB and FeCl$_2$.

In their research using FeCl$_2$ as activating agents to recycle tomato stem waste, Fu et al. (2017) reported that Fe$_2$O$_3$ from FeCl$_2$ decomposition could be reduced by
charred organics to produce Fe$_3$O$_4$ or even Fe$^0$ (reactions (3) and (4))

\[ > 490°C: 3\text{Fe}_2\text{O}_3 + \text{C} \rightarrow 2\text{Fe}_3\text{O}_4 + \text{CO} \]  
\[ 700 – 800°C: \text{Fe}_3\text{O}_4 + \text{C} \rightarrow 3\text{Fe}^0 + 2\text{CO}_2 \]  

(3)  
(4)

It's possible the higher mass loss from WPCB/FeCl$_2$ was the result of reduction of Fe$_2$O$_3$ by carbon to Fe$_3$O$_4$ or Fe$^0$ in WPCB.

XRD, FTIR patterns and SEM images of WPCB/FeCl$_2$ residual solids were shown in Fig.3(b)-(d). XRD pattern (Fig.3(b)) showed Cu crystals in the raw WPCB. Fe$_2$O$_3$ was the main crystals found in thermal treatment of FeCl$_2$ at 650°C. FeOOH was also detected at 2θ of 28.9º. The results here were in accordance with those indicated in reactions (1) and (2). Fe$_3$O$_4$ was observed for WPCB/FeCl$_2$ mixture treated at 450°C and 650°C together with Fe$_2$O$_3$. This proved that Fe$_2$O$_3$ from FeCl$_2$ decomposition was reduced to form Fe$_3$O$_4$ when WPCB was present. Also, peaks at 2θ of 53º and 57º were believed to be those of FeBr$_3$. This means that Br from WPCB was fixed by FeCl$_2$ decomposition products. However, the absence of FeBr$_3$ in XRD patterns of 450°C did not indicate no Br-fixing occurred. It only means the resultant FeBr$_3$ may not be crystalline.

Fig. 3(c) were the FTIR patterns. Peaks around 3440-3407 cm$^{-1}$ were assigned to the vibration of O-H stretching (Ng et al. 2009). Bands at about 1400 -1600 cm$^{-1}$ were the C=C vibration in the benzene ring skeleton while those at 1040 cm$^{-1}$ were attributed to the C-O stretching (Zhang et al. 2015). In addition, C-Br peaks were found in all samples at 459-471 cm$^{-1}$ as brominated epoxy was commonly employed in WPCB as fire retardant while Fe-O peaks at 603-620 cm$^{-1}$ were found only in WPCB/FeCl$_2$.
pyrolyzed samples. Fig. 3(d) is the SEM image of WPCB/FeCl₂ pyrolyzed at 650°C (WPCB/FeCl₂-650). WPCB consist mainly of cylindrical glass fibers and small particles. After pyrolysis with FeCl₂, glass fibers disintegrated due to the corrosive effects of HCl from FeCl₂ decomposition.

### 3.3 Pyrolysis with FeCl₃

Fig. 4 is the mass loss, XRD, FTIR patterns and SEM image of WPCB pyrolyzed with FeCl₃.

FeCl₃ was converted to FeOOH under high temperature (reactions (5)) which could then undergo further reactions such as reactions (1) and (2) above to form other iron compounds.
262 oxides (Xu et al. 2019).

263 \[ 200 - 330^\circ \text{C}: \text{FeCl}_3 + 2\text{H}_2\text{O} \rightarrow \text{FeOCl} \cdot \text{H}_2\text{O} + 2\text{HCl} \rightarrow \text{FeOOH} + 3\text{HCl} \] \hspace{1cm} (5)

264 According to mass loss results in Fig. 4(a), FeCl\(_3\) suffered continuous mass loss as temperature rose from 350°C to 650°C then plateaued. 55% of the original FeCl\(_3\) was lost during thermal treatment. Mass losses of WPCB/FeCl\(_3\) mixture varied from 20% to about 39% which was also slightly higher than the calculated mass losses. This could also be due to the interaction between iron oxides and carbon produced from WPCB pyrolysis as shown in reactions (3) and (4).

269 The reduction of Fe\(_2\)O\(_3\) to Fe\(_3\)O\(_4\) was proven by XRD patterns in Fig.4(b). Fe\(_3\)O\(_4\) were detected in WPCB/FeCl\(_3\) pyrolyzed at 350°C and 550°C. As comparison, the dominant crystalline phases detected in FeCl\(_3\) treated under N\(_2\) at 550°C were FeOOH and Fe\(_2\)O\(_3\). Peaks at 2\( \theta \) of 24.8º and 48.9º were ascribed to FeOCl\( \cdot \)H\(_2\)O. These were found in pyrolysis samples of FeCl\(_3\) and WPCB/FeCl\(_3\) treated at 350°C but not samples treated at 550°C. This means that FeOCl\( \cdot \)H\(_2\)O was an intermediate during decomposition which was converted at high temperature. In addition, FeBr\(_3\) crystals were found in the solid residuals treated at temperature 550°C probably due to reaction between Br from WPCB and iron oxides.

279 Fig. 4(c) were the FTIR patterns of original WPCB and two pyrolyzed WPCB/FeCl\(_3\) samples. Peaks of O-H, C=C, C-O and C-Br structures were detected in all samples together with Fe-O structures. SEM image of WPCB/FeCl\(_3\) treated at 550°C showed glass fibers in cylindrical shapes but cavities were observed on the surface of glass giber indicating also that the addition of FeCl\(_3\) led to the corrosion of glass fiber.
3.4 Pyrolysis with FeSO₄

Fig. 5 is the mass loss, XRD, FTIR and SEM analysis results WPCB pyrolyzed with FeSO₄.

![Graph showing mass loss, XRD, FTIR, and SEM analysis results for WPCB pyrolyzed with FeSO₄.]

In their study on thermal decomposition of FeSO₄·6H₂O, Masset et al. (2006) revealed that FeSO₄·6H₂O dehydrated step-wise from FeSO₄·6H₂O to FeSO₄·4H₂O to FeSO₄·H₂O at temperature from 70-310°C. Decomposition of FeSO₄ to hematite (Fe₂O₃) occurred at high temperature (reactions (6) and (7)). FeSO₄ completely converted to Fe₂(SO₄)₃ and Fe₂O₃ at 550°C and Fe₂(SO₄)₃ was found to be the only intermediate compound during the decomposition. The mass loss by FeSO₄ under N₂
shown in Fig. 5(a) in this research correlated well with Masset et al.’s findings. Sharp mass loss of FeSO₄ was recorded only when pyrolysis temperature was greater than 550°C.

$$\begin{align*}
525 - 650°C: 6\text{FeSO}_4 &\rightarrow \text{Fe}_2(\text{SO}_4)_3 + 2\text{Fe}_2\text{O}_3 + 2\text{SO}_2 & (6)
\end{align*}$$

$$\begin{align*}
527 - 710°C: 2\text{FeSO}_4 &\rightarrow \text{Fe}_2\text{O}_3 + 2\text{SO}_2 + \text{O}_2 & (7)
\end{align*}$$

X-ray diffraction was conducted to identify the crystalline phases formed during pyrolysis. XRD pattern of FeSO₄ treated at 550°C for 1 h in Fig. 5(b) showed the presence of both Fe₂(SO₄)₃ and Fe₂O₃. These two phases were also detected with WPCB/FeSO₄ treated at 350°C. In contrast, Fe₂(SO₄)₃ was not found in WPCB/FeSO₄ treated at 550°C indicating a complete conversion to Fe₂O₃ at high temperature. FeBr₃ was present at WPCB co-pyrolyzed with FeSO₄ at 550°C but not at sample of 350°C. FTIR pattern were quite similar to those co-pyrolyzed with FeCl₂ and FeCl₃. Peaks at 3420-3450 cm⁻¹, 1510 cm⁻¹, 1040-1070 cm⁻¹, 460-480 cm⁻¹ were due to the vibration of O-H, C=C, C-O and C-Br structures respectively. Fe-O vibrations were detected around 590 cm⁻¹. SEM image of WPCB/FeSO₄ treated at 550°C showed glass fibers remaining cylindrical with very little disintegration. The corrosive effects of FeSO₄ was much less than those of FeCl₂ and FeCl₃.

Overall, co-pyrolysis of WPCB with FeCl₂, FeCl₃ and FeSO₄ generated solids with different morphology and composition. Table 1 further lists the elemental composition, pore volume and surface areas of these solids. About half of the carbon in raw WPCB was lost during pyrolysis. The addition of FeSO₄ had resulted in obvious increase in S content. Solid manifested certain porosity with surface areas about 20.6-26.5 m²/g.
Table 1  elemental analysis, total pore volumes and surface areas of pyrolyzed solids

| Sample            | Elemental Analysis (%) | Total pore volume (cm$^3$/g) | BET Surface area (m$^2$/g) |
|-------------------|------------------------|------------------------------|-----------------------------|
|                   | C  | H  | N  | S  |               |                            |
| WPCB              | 24.13 | 1.95 | 0.32 | 0.31 | -                  | -                          |
| WPCB/FeCl$\text{2}$-650 | 12.80 | 2.26 | 0.28 | 0.37 | 0.12               | 23.4                       |
| WPCB/FeCl$\text{3}$-550 | 14.34 | 1.60 | 0.35 | 0.36 | 0.089              | 20.6                       |
| WPCB/FeSO$\text{4}$-550 | 11.60 | 0.47 | 0.20 | 5.63 | 0.12               | 26.5                       |

Pyrolysis changed physical and chemical properties of WPCB. Based on liquid yield, it is optimal to conduct pyrolysis at mass ratio of 0.5 and temperatures of 550-650°C. Part of Br was fixed as FeBr$_3$ crystals at these temperatures. The resultant solids consist of chars, iron oxides and metals from raw WPCB and added reagents. Research on recovering of residual metals (e.g. Cu), toxicity leaching and potential reuse of the pyrolysis solids are currently underway. It could be a potential source of building materials. Gao et al. (2019) used pyrolysis products as carbon and iron source to synthesize magnetic carbon fibers.

3.5 Characterization of liquid and gas products

In addition to the characterization of solids, the composition of liquid products was determined by GC-MS as shown in Fig.6. Liquids were collected from pyrolysis at mass ratio of 0.5 and temperature of 650°C for WPCB/FeCl$_2$ and 550°C for WPCB, WPCB/FeCl$_3$ and WPCB/FeSO$_4$. Phenol and its derivatives were identified as the main products. Monitoring of liquid composition showed varied presence of brominated organics with the addition of co-pyrolysis reagent. Pyrolysis of WPCB generated a wealth of brominated organics including 2,4-dibromophenol and 2,4,6-tribromophenol,
2-bromo-4-methylphenol. In contrast, liquid products of FeCl$_2$ were dominated by phenol and no brominated organics compounds were detected. Chloride and bromide were released as HCl and HBr into the liquid phase as shown in Fig. 6(b). Brominated organics were also found when FeCl$_3$ and FeSO$_4$ were used. Particularly, co-pyrolysis with FeCl$_3$ actually led to the generation of chlorinated compound, i.e. 4-chloro-2-(1-methylethyl)phenol in Fig.6(c).

Fig.6 GC-MS results for pyrolysis liquid from (a) WPCB-550; (b) WPCB/FeCl$_2$-650; (c) WPCB/FeCl$_3$-550; (d) WPCB/FeSO$_4$-550

Compared with FeCl$_3$ and FeSO$_4$, conversion of FeCl$_2$ to FeOOH and HCl started at lower temperature (120-200°C) according to reaction (1). It is possible that the lack
of chlorinated products was due to the fact that Cl in FeCl$_2$ was quickly released at low
temperature as HCl which prevented the formation of chlorinated products. At the same
time, iron oxides thus generated acted as Br-fixing agent when Br-containing fire
retardant started to degrade. Also, the addition of FeCl$_2$ suppressed the bromination of
organics as Br released from fire retardant was either released as HBr or fixed as FeBr$_3$.
The mechanism of Br-fixing will be further clarified by studying the mass balance of
Br in future research. Based on the quantity and quality of liquid products, FeCl$_2$ could
be recommended as a co-pyrolysis reagent.

In addition, pyrolysis exhaust was collected and analyzed. CO$_2$, CH$_4$ and H$_2$ were
the main compounds detected. It appeared that no toxic gas was generated during
pyrolysis.

4. Conclusions

WPCB was co-pyrolyzed with Fe$_3$O$_4$, FeCl$_2$, FeCl$_3$ and FeSO$_4$. Liquid yield was
used as the screening factor which showed that the addition of FeCl$_2$, FeCl$_3$ and FeSO$_4$
was conducive to liquid production. The optimal mass ratio and temperatures were 0.5
and 550-650°C for liquid yields which ranged from 18-20%. About half of the C in raw
WPCB remained in solid phase and was present as chars. Iron compounds were
decomposed to iron oxides at high temperature. Characterization of liquid by GC-MS
revealed that phenol and its homologues are the main products which could be a
valuable chemical source. FeCl$_2$, FeCl$_3$ and FeSO$_4$ showed different Br-fixing capacity.
No brominated or chlorinated organics were found in liquid products from co-pyrolysis
with FeCl$_2$. The addition of FeCl$_2$ suppressed the generation of brominated organics.
Br was either fixed as FeBr$_3$ in solids or released as HBr to liquid. In summary, the roles of co-pyrolysis reagent such as FeCl$_2$ were two-fold. First, it could improve the decomposition of macro-molecule in WPCB and increase liquid yield. Secondly, it reduced the bromination of organics and prevent excessive release of brominated organics.

**Declarations**

- Ethics approval and consent to participate
  
  Not applicable

- Consent for publication
  
  Not applicable

- Availability of data and materials
  
  All data generated or analyzed during this study are included in this published article

- Author contributions
  
  All authors contributed to the study conception and design. Material preparation, data collection and analysis were performed by Weifang Chen, Yongkai Shu, Yonglun Li, Yanjun Chen and Jianbo Wei. The first draft of the manuscript was written by Weifang Chen and Yongkai Shu. All authors commented on previous versions of the manuscript. All authors read and approved the final manuscript.

- Competing interests
  
  The authors declare that they have no competing interests

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Figures

Figure 1

SEM image of the original WPCB

(a) Liquid Yield (%)  
- WPCB  
- Fe3O4  
- FeCl2  
- FeCl3  
- FeSO4

(b) Liquid yield (%)  
- Fe3O4  
- FeCl2  
- FeCl3  
- FeSO4

Figure 2
Liquid yield of WPCB with different iron compounds with the change of (a) temperature and (b) mass ratio

Figure 3

Characterization of WPCB, FeCl2 and WPCB/FeCl2: (a) mass loss; (b) XRD; (c) FTIR; (d) SEM image of WPCB/FeCl2-650
Figure 4

Characterization of WPCB, FeCl3 and WPCB/FeCl3: (a) mass loss; (b) XRD; (c) FTIR; (d) SEM image of WPCB/FeCl3-550
Figure 5

Characterization of WPCB, FeSO4 and WPCB/FeSO4: (a) mass loss; (b) XRD; (c) FTIR; (d) SEM image of WPCB/FeSO4-550
Figure 6

GC-MS results for pyrolysis liquid from (a) WPCB-550; (b) WPCB/FeCl₂-650; (c) WPCB/FeCl₃-550; (d) WPCB/FeSO₄-550