Environmental Life Cycle Assessment of Ammonia-Based Electricity

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1. Introduction

Ammonia (NH₃) is currently gaining considerable interest as a zero-carbon energy vector from academic institutions, governmental bodies, and industrial companies. The potential use of ammonia in power generation systems such as fuel cells, internal combustion engines, furnaces, and gas turbines is under scrutiny at different power scales [1], with considerable efforts to tackle emissions and enable reliable, stable processes [2]. Progress on the subject has led to international programs that aim to develop innovative technologies to demonstrate ammonia’s use to store renewable energy from wind, solar, and even marine resources [3]. However, ammonia as fuel still requires further research, especially when demonstrating an integrated approach that considers all the impacts in the production and utilization chain of ammonia produced with different methods and in various regions.
Therefore, this work attempts to bring all these points under a life cycle assessment umbrella to demonstrate the potential impact of integrated ammonia-based units fueled from ammonia produced in different parts of the world.

The challenge of considering ammonia as an energy carrier does not only lie in the technical complexities of using a slow-reacting molecule with the potential of NOX formation [4] but also in the economic and logistical aspects of the concept behind generating green ammonia via renewable sources. The challenges of implementing large-scale green ammonia production and its novel uses must be identified and overcome if its development and deployment are to be accelerated and fully capitalized on [5]. Overall, some of these challenges include upfront investment costs (i.e., EUR 300–350 tNH$_3$ compared to EUR 440–630 tNH$_3$ via green ammonia [6]), existing capacity, market uncertainty (i.e., use of ammonia for marine applications may vary from 25% to 99% by 2050 [7,8]), erosion of revenues due to other vectors, government support, and last but not least, environmental impacts of the “hydrogen through ammonia” economy concept. Hence, current efforts focus on developing studies and demonstration systems capable of meeting these challenges. As for environmental impacts, the use of novel tools has improved insight into the effects of various ammonia-based systems, primarily via life cycle assessments.

A life cycle assessment (LCA) is defined as the systematic analysis of the potential environmental impacts of products or services during their entire life cycle. Ammonia has been under the scope of LCA analysts for several years. Wei and En-ke [9] performed an analysis of the use of coal-based ammonia for vehicular applications, finding the concept feasible. Similarly, Bicer et al. have approached the subject by evaluating different ammonia production methods [10], aerospace, marine, and vehicle applications [11–13], and integration of concepts [14]. Their results denote the impacts on eutrophication, abiotic depletion, global warming potential, and human health. Although the results are mixed and some options are better than others, it is evident from all their work that ammonia can be obtained with low environmental impacts, while the use of current technologies can enable similar or lesser effects on health and the environment if ammonia replaces fossil-based fuels. Cox and Treyer [15] also conducted assessments for the use of fuel cells that could replace standard power systems. Their study focused on alkaline units, denoting that there is still work to do to improve cell lifetime, power densities, and fuel consumption for less impactful devices. Further work was presented by Razon and Valera-Medina [16], who studied ammonia in gas turbines. Blends of ammonia–methane were evaluated, with results showing that ammonia produced from nuclear or wind energy could potentially support global decarbonisation with a −27% to −32% reduction in global warming potential. Table 1 summarizes these studies that analyze the environmental profile of ammonia-based electricity.

Table 1. Life cycle assessment studies of ammonia-based electricity.

| Reference          | System                      | Objective/Goal                                           | System Boundaries                                                                 | Functional Unit | Impact Categories                                      |
|--------------------|-----------------------------|----------------------------------------------------------|----------------------------------------------------------------------------------|-----------------|--------------------------------------------------------|
| Bicer and Dincer [12] | Ammonia utilization in transportation and power generation. | To study the environmental impact of ammonia as fuel in passenger cars in cities and generation in a gas turbine power plant. | Manufacturing, operation, maintenance, and disposal of the vehicle. Construction of power plant, ammonia production (wind-based), transportation of ammonia, and power plant operation. | 1 traveled km and 1 MJ energy generated. | Abiotic depletion potential, acidification potential, global warming potential, ozone layer depletion potential. |
Most studies conclude that current technologies used to produce green ammonia need to be improved. Pfromm [18] concludes in his assessment that using the available industrial-scale technology, the energy demand for a process based on electrical energy to provide hydrogen and nitrogen is about 14% higher than that of the conventional natural gas-based process. On the other hand, according to Smith [19], electrochemical-based ammonia could achieve reductions in energy consumption and CO₂ emissions by 50% and 78%, respectively, in comparison with methane-based ammonia. Hence, technological developments are needed in conjunction with better cycles that can increase the overall efficiency of ammonia power systems.

On that line of research, ammonia cycles have also experienced an increase in analysis. Relevant work presented by Keller et al. [20] denotes how the use of ammonia could potentially be employed with efficient profiles, while by-products of the process (mainly water, nitrogen, and traces of hydrogen that are burned in a heat recovery steam generator) can boost the final thermal efficiency, keeping NOₓ emissions low. Further work performed by Mashruk et al. [21] approached the use of ammonia via hydrogen and humidified blends. The results show that using ammonia with hydrogen/steam blends potentially can reduce NOₓ in an order of magnitude, while combustion efficiencies remain high with the implementation of two-stage combustion systems. The works were expanded by Gutesa-Bozo et al. [22], who presented an innovative cycle that uses ammonia for power generation and cooling and heating purposes. The results show that efficiencies of up to
60% can be achieved using ammonia blends. The results are in accord with those found by other studies [23]. Hence, that study was used as a background in this paper, combining an efficient cooling/heating/power cycle with state-of-the-art ammonia production methods in a detailed LCA that is critically needed to progress to a physical demonstration. The work is expanded to different regions, showing how ammonia production can have different impacts via other methods and resources.

Ammonia production is usually centralized, but distributed production is required when coping with the geographic isolation and intermittency of renewable energy [19]. Several researchers have addressed the environmental impact of ammonia production considering different scales, feedstocks, and pathways; however, few of them include a description of ammonia life cycle inventories at different production scales. Moreover, an environmental evaluation of a co-generation heat and power (CHP) system based on ammonia has not been analyzed from a life cycle perspective in the literature.

Ammonia presents several barriers before it can be fully deployed as an energy carrier at a global scale [5]. First, ammonia needs to be produced in a simple, cheap manner. Current methods still rely on the production of hydrogen via steam-methane reforming, hence heavily contributing to climate change. Unfortunately, the use of electrolyzers (as in this work) still poses a heavy economic burden for green ammonia, a subject that is under scrutiny worldwide. Second, utility-scale power generation needs to be addressed with efficient systems that mitigate the production of unwanted emissions such as nitrogen oxides. Nitrous oxide (N₂O) is a strong greenhouse gas that can be produced from ammonia combustion systems, thus reducing the eco-friendly nature of using ammonia as a carbon-fuel replacement if these systems are not properly designed, which is another state-of-the-art subject under continuous research. Third, public perception needs to be fully acknowledged. Understanding how communities and individual users can use ammonia as a carbon-free fuel is still poorly understood. Bad perceptions of the chemical can stop stakeholders’ interest, funding, and progress on the topic. Finally, all the developments need to be profitable with healthy economic profiles, otherwise companies of all sizes will not invest in ammonia technologies. Therefore, this work approaches the latter two barriers with a study that supports a better understanding of the impacts of using ammonia as an energy vector for a better acknowledgement of its potential application in different locations, while the results also serve as the basis for future economic analyses.

The objectives of this study are: (i) to present a life cycle inventory of ammonia production in small (20,000 t/year) and medium (100,000 t/year) scale plants; (ii) to compare the environmental performance of ammonia production using different energy resources and different technology scales; and (iii) to compare the environmental performance of ammonia-based electricity generated in the studied CHP system using the different ammonia production pathways. This research will serve as a basis to understand and discuss the role of ammonia through different commercially available routes to generate electricity as support for decarbonization.

2. Materials and Methods

This work aims to evaluate the environmental profile of ammonia as an energy carrier, following the methodological framework provided in the ISO Standards 14040:2006 and 14044:2006 [24,25]. LCA has been used successfully to evaluate the environmental sustainability of electricity generation systems [26–28].

A cradle-to-gate assessment was developed for both studied systems. Figure 1 presents the streamlined ammonia systems for natural gas steam reforming (a) and water electrolysis (b). Figure 2 illustrates the whole system, including extracting raw materials and energy resources, processing, manufacturing, and distributing ammonia, CHP plant operation, plant construction, and displaced heat from natural gas.
Figure 1. System boundaries of the ammonia production system: (a) based on natural gas via steam methane reforming to obtain hydrogen; (b) based on water via electrolysis to obtain hydrogen.

Figure 2. The system boundary of the ammonia-based electricity generation system. The CHP has a thermal output of 2.58 kWh per 1 kWh of electricity generated, which can be used as industrial or district heating. This heat was integrated into the model as a displacement of heat from natural gas. * Natural gas is the only input in the production of ammonia via SMR. The quantitative values of materials and energy flows are described in detail in Tables S1 to S3 of the Supplementary Materials.
The functional unit, FU, is the quantity in which all the inputs and outputs are calculated; as well, it serves as a basis to estimate the environmental impacts and compare with other systems with the same function. Therefore, two FU are considered in this work: “1 ton of ammonia (NH$_3$) at the gate of the production facility” and “1 kWh of electricity at the CHP plant”.

The life cycle inventory (LCI) stage is the systematization of the input and output flows for the product system. The LCI of the ammonia production system was developed based on primary data from Proton Ventures BV [29] and supplemented with secondary data from the literature [30,31]. The estimation of fugitive emissions during the storage and transport stage was based on the work of Al-Breiki and Bicer [32,33]. The power plant model was based on the operation cycle proposed by Guteša Božo et al. [22] that uses a mix of humidified ammonia/hydrogen as fuel (NH$_3$ is partially cracked to produce a 30/70 H$_2$/NH$_3$ blend). Figure 3 shows a representation of the power cycle. The LCI in all the system stages was completed with background data obtained from the Ecoinvent database 3.7.1 [34]. Background data include material and energy flows needed to extract, process, manufacture, and distribute fuels, electricity, products, and capital goods.

Various scenarios were studied to analyze the environmental feasibility of synthesized ammonia as a carrier in the global energy market. Ammonia can be produced from nitrogen (N$_2$) and hydrogen (H$_2$) through the Haber–Bosch process. Purified nitrogen is obtained from the atmospheric air via an air separation unit (ASU), which requires a certain amount of energy. Depending on the location and production scale, hydrogen gas can be obtained from different feedstocks (methane, water, biomass) and various energy sources (fossil resources, electricity from renewables, or nuclear). Steam methane reforming (SMR) is the most used technology to produce hydrogen (around 80% of the global production of ammonia [35]).

In contrast, using an electrolyzer with water as feedstock and electricity as an energy source is gaining more relevance nowadays [3,5]. Hydrogen production represents most of the energy requirement, typically more than 90% of the total energy requirement [36]. Both hydrogen production technologies were included in this study, i.e., steam methane reforming and electrolysis.

Regarding locations, the UK is the baseline scenario where the co-generation power plant is situated. Considering the future role that ammonia could have as a storage carrier in a decarbonized global energy market [10], other locations outside the UK were considered to produce green ammonia, potentially for import of green ammonia to the UK.
A variety of research shows promising technical and economic feasibility of producing hydrogen/ammonia from renewable sources and exporting to other locations in Europe or Asia [37–39]. In this regard, Australia [40,41], Morocco [42,43], and Chile [44,45] are in the lead with the development of green ammonia production facilities. Brazil, one of the major consumers of fertilizers that also has a significant renewable energy capacity, and Iceland, were also considered for this study. The United Arab Emirates are expanding their energy portfolio, with nuclear [46] and ammonia [47,48]; therefore, a scenario considering ammonia production based on nuclear energy from UAE was included.

Ammonia production is mainly centralized, but distributed production may be required in isolated areas. Distributed ammonia production opens up transformative opportunities for applications besides fertilizers. For this, the Haber–Bosch process should be adapted to the geographic isolation and intermittency of renewable energy sources, decarbonized, be made more energy efficient, and deployed in low capitol projects [19]. Indeed, the production scale is an issue as well: when scaling down, heat losses and energy consumption increase [36]. For both gray (SMR) and green (from renewable energy) ammonia, a small-scale scenario of ammonia production (with a yield of 20,000 t per annum) was modeled for the UK. Medium-scale production (100,000 t per annum) scenarios for production of gray, green, and blue ammonia (SMR with carbon capture and storage—CCS) are modeled for the UK and the other locations to later import to the UK. Table 2 includes a description of the scenarios under study.

| Scenario Code | Description |
|---------------|-------------|
| SMR-S-UK      | Methane to ammonia via steam methane reforming process, small capacity (20,000 t of NH₃ per annum) in the UK. |
| SMR-M-UK      | Methane to ammonia via steam methane reforming process, mid-capacity (100,000 t of NH₃ per annum) in the UK. |
| SMR-CCS-M-UK  | Methane to ammonia via steam methane reforming process, with carbon capture and storage, mid-capacity (100,000 t of NH₃ per annum) in the UK. |
| E-W-S-UK      | Power to ammonia via electrolysis process, electricity from wind, small capacity (20,000 t of NH₃ per annum) in the UK. |
| E-W-M-UK      | Power to ammonia via electrolysis process, electricity from wind, mid-capacity (100,000 t of NH₃ per annum) in the UK. |
| E-W-M-MA      | Power to ammonia via electrolysis process, electricity from wind, mid-capacity (100,000 t of NH₃ per annum) in Morocco. |
| E-W-M-AU      | Power to ammonia via electrolysis process, electricity from wind, mid-capacity (100,000 t of NH₃ per annum) in Australia. |
| E-PV-M-CL     | Power to ammonia via electrolysis process, electricity from photovoltaics, mid-capacity (100,000 t of NH₃ per annum) in Chile. |
| E-H-M-BR      | Power to ammonia via electrolysis process, electricity from hydropower, mid-capacity (100,000 t of NH₃ per annum) in Brazil. |
| E-GT-M-IS     | Power to ammonia via electrolysis process, electricity from geothermal, mid-capacity (100,000 t of NH₃ per annum) in Iceland. |
| E-N-M-AE      | Power to ammonia via electrolysis process, electricity from nuclear, mid-capacity (100,000 t of NH₃ per annum) in the United Arab Emirates. |

SMR: steam methane reforming, E: electrolysis, CCS: carbon capture and storage, S: small-scale 20,000 ton per annum ammonia plant, M: mid-scale 100,000 ton per annum ammonia plant, PV: photovoltaic, W: wind, H: hydro, GT: geothermal, N: nuclear, UK: United Kingdom, MA: Morocco, AU: Australia, CL: Chile, BR: Brazil, IS: Iceland, AE: United Arab Emirates.

Ammonia synthesis is a highly energy-intensive process; therefore, the energy efficiency of its production is an essential factor to consider when evaluating the sustainability of large-scale production. This study addresses this issue by comparing the future environmental profile of green ammonia, considering future trends in electrolyzer technology [19].
Scenarios using the country electricity generation mix and a conventional natural gas power plant to assess the environmental cost of producing ammonia with non-renewable sources were included. Natural gas is used as this is likely the marginal energy technology used in the countries where ammonia is produced. The marginal energy technology represents a case where the generation technology must cover a new demand on the margin of that energy market.

The life cycle impact assessment (LCIA) was performed using the characterization factors comprised in the ReCiPe2016 Midpoint(H) v1.13 methodology [49] and was modeled using the OpenLCA software [50]. Impact categories included in the analysis are listed in Table 3.

### Table 3. Impact categories included in the LCA (ReCiPe2016 Midpoint(H) [49]).

| Impact Category                  | Indicator                          | Characterisation Factor                          | Unit    |
|----------------------------------|------------------------------------|--------------------------------------------------|---------|
| Climate change                   | Infra-red radiative forcing increase| Global warming potential—GWP100                   | kg CO₂-eq |
| Fossil resource scarcity         | Upper heating value                | Fossil depletion potential—FDP                   | kg oil-eq |
| Freshwater eutrophication        | Phosphorus increase in freshwater  | Freshwater eutrophication potential—FEP          | kg P-eq  |
| Ozone depletion                  | Stratospheric ozone decrease       | Ozone depletion potential—ODPinf                 | kg CFC-11-eq |
| Photochemical oxidant formation  | Tropospheric ozone increase        | Photochemical oxidant formation potential—POFP   | kg NMVOC-eq |
| Terrestrial acidification         | Proton increase in natural soils   | Terrestrial acidification potential—TAP100       | kg SO₂-eq |
| Ionising radiation               | Absorbed dose increase             | Ionizing radiation potential—IRP                 | kg U235-eq |

In addition, the ReCiPe2016 Midpoint(I) methodology [49] was used to assess the potential reduction in GHG emissions between gray (methane-based) and blue (methane-based with CCS) ammonia.

### 3. Results and Discussion

Supplementary Material shows the life cycle inventory of the ammonia production system (Table S1), the storage and transport of ammonia (Table S2), and the electricity generation production system (Table S3).

#### 3.1. Life Cycle Impact Assessment of Ammonia Production

Figure 4 shows the impact characterization of the ammonia production system and the contribution of the main processes relative to each impact category. In the present study, most impacts of the electrolysis based-ammonia scenarios are related to the electrolysis process. This agrees with the literature, as the effect of ammonia production depends mainly on the methods used to produce hydrogen [51]; indeed, hydrogen synthesis accounts for around 90% of the carbon emissions for ammonia synthesis [35].

All the LCA studies include mainly GHG emissions metrics (i.e., global warming potential, carbon footprint, climate change). However, few include additional impact categories under different methods, such as CML 2001 [10], Eco-Indicator 99, or ReCiPe [52].
Figure 4. Contribution analysis for the impact category indicator results of 1 kg ammonia at the production plant gate for (a) global warming potential (GWP100), (b) fossil depletion potential (FDP), (c) freshwater eutrophication potential (FEP), (d) ozone depletion potential (ODP), (e) photochemical oxidation formation potential (POFP), (f) terrestrial acidification potential TAP100, (g) ionizing radiation potential (IRP). SMR: steam methane reforming, E: electrolysis, CCS: carbon capture and storage, S: small-scale 20,000 ton per annum ammonia plant, M: mid-scale 100,000 ton per annum ammonia plant, PV: photovoltaic, W: wind, H: hydro, GT: geothermal, N: nuclear, UK: United Kingdom, MA: Morocco, AU: Australia, CL: Chile, BR: Brazil, IS: Iceland, AE: United Arab Emirates.
The results show that SMR production pathways have the highest GWP values in general: 2.75 t CO\textsubscript{2}-eq/t NH\textsubscript{3} for mid-scale production, 2.6 t CO\textsubscript{2}-eq/t NH\textsubscript{3} for small-scale. This is due to the use of fossil resource (natural gas) as feedstock and fuel, and thus it is also related to the higher fossil depletion potential (FDP) related to SMR processes. The results for the other scenarios (i.e., all electrolysis based on renewable energy scenarios and CCS scenario) are below 0.7 t CO\textsubscript{2}-eq/t NH\textsubscript{3}. For the SMR scenarios, the natural gas extraction and transport is the most relevant process for the FDP: it is related to 94% of FDP in mid-scale production and 84% in small scale. These results are in line with reviewed studies. Ghavam et al. [53] obtained 2.16 t CO\textsubscript{2}-eq/t NH\textsubscript{3}. At the same time, Al-Breiki and Bicer [32] found that ammonia production via steam methane reforming with and without carbon capture and storage (CCS) emits 1.68 and 2.24 t CO\textsubscript{2}-eq/t NH\textsubscript{3}, respectively.

Adding carbon capture processes to steam reforming improves the performance in terms of carbon emissions. In the present study, when adding CCS, GHG emissions decreased by almost 60%. At the same time, other impact categories were raised (FEP increased by 56%, IRP increased by 47%, others increased less than 8%; this is related to the extra energy needed to run the CCS system). According to Tock [54], there is a small increment in energy consumption due to CCS compressing the carbon dioxide since the CO\textsubscript{2} is already separated to purify the H\textsubscript{2} in the SMR. However, Dufour et al. [55] found that CCS worsens other environmental impact categories (single score Eco-indicator 95 mPt: GHGs decreased by two-thirds, acidification increased by around half, heavy metals increased by around one-third, winter smog increased by half). Nevertheless, this depends on the energy mix or electricity source.

Hydrogen production is predominant in the environmental profile of ammonia production, especially for the impacts of climate change, fossil depletion, and ozone layer depletion. There are mixed results for blue hydrogen (fossil-based hydrogen with CCS) regarding global warming potential in the literature. Atitlan et al. [56] reported a range from 40 to 90 g CO\textsubscript{2}-eq per MJ of L-H\textsubscript{2}. In contrast, Howarth and Jacobson [57] estimated 135 or 139 g CO\textsubscript{2}-eq per MJ with and without flue-gas capture, respectively, considering a time horizon of 20 years for this indicator. A sensitivity analysis for a 20-year time frame was performed. The results show (Figure 5) that the implementation of CCS entails a reduction of 44% of the carbon footprint of 1 t ammonia via SMR, which is less than previously estimated for GWP100.

![Figure 5. Global warming potential (GWP) considering 20- and 100-year time horizons for 1 t of NH\textsubscript{3} via steam methane reforming (SMR) with and without carbon capture and storage (CCS).](image)

This study shows that renewable energy electrolysis-based ammonia has a favorable environmental performance in terms of climate change. Wind-based electrolysis results are between 0.24 and 0.54 t CO\textsubscript{2}-eq/t NH\textsubscript{3}. Singh et al. [51] found that ammonia from wind emits 0.496 t CO\textsubscript{2}-eq/t NH\textsubscript{3}. In this study, 1 t of hydropower electrolysis-based ammonia emits 0.66 t CO\textsubscript{2}-eq, solar PV electrolysis-based ammonia resulted in 0.7 t CO\textsubscript{2}-eq, and geothermal electrolysis-based ammonia resulted in 0.27 t CO\textsubscript{2}-eq. In previous studies, it
was estimated to be 0.38 t CO\(_2\)-eq per t of hydropower-based ammonia [10] and 1.28 t CO\(_2\)-eq per t of solar PV ammonia [51]. The GHG emissions are mostly related to the materials and energy needed to build and install the power plant facilities.

Nuclear-based ammonia (pink ammonia) production has the lowest GWP (0.09 t CO\(_2\)-eq/t NH\(_3\)). However, the ionizing radiation potential indicator performs very unfavorably (around 7000 kg U\(_{235}\)-eq/t NH\(_3\) against less than 300 in the other scenarios). Likewise, Bicer and Dincer [58] found that nuclear electrolysis-based ammonia has lower CO\(_2\) equivalent emissions than SMR, generating between 0.45 and 0.6 t CO\(_2\)-eq/t NH\(_3\). Still, it worsens the environmental performance in terms of human toxicity. Bicer et al. [10] reported 0.84 t CO\(_2\)-eq/t NH\(_3\), and Karaca et al. [59] reported GWP for different nuclear electrolysis-based ammonia production ranging from 0.18 to 0.337 t CO\(_2\)-eq/t NH\(_3\).

3.2. Life Cycle Impact Assessment of Ammonia-Based Electricity

Figure 6 shows the impact characterization of 1 kWh of electricity based on ammonia. Results show that ammonia production is the main contributor to the environmental impacts of ammonia-based electricity and that the use of surplus heat for district or industrial applications offsets some of the environmental burdens in all impact categories.

Displacement of heat from natural gas for utilization in industrial or domestic applications reduces the carbon footprint of each kWh of electricity-based ammonia in the range of 30% to 40%; therefore, the negative values obtained imply an overall positive impact on the environment. Without considering co-generation, the eight electrolysis-based ammonia scenarios show an advantage over methane-based ammonia, even being below the UK’s grid mix (0.325 kg CO\(_2\)-eq/kWh) or a conventional gas turbine power plant in the UK (0.511 kg CO\(_2\)-eq/kWh); values taken from Ecoinvent [34]. Regarding the emissions during transport, the results show that for locations such as Brazil, Chile, UAE, and Australia (distances greater than 3000 nautical miles), the contribution of the emissions during transport is significant (between 7% for hydro from Brazil to 39% for nuclear from UAE). The share contribution of transport depends on the distance and the overall GHG; hence, medium and long distances, combined with the lowest global GHG emissions, resulted in a high percentage contribution from transport emissions. Likewise, the results of Al-Breiki and Bicer [32] show that for 10,000 nmi, the share of carbon emissions during transport is 3% for NH\(_3\) from natural gas without CCS and 15% for NH\(_3\) from wind electrolysis, while for 20,000 nmi, the share increases to 7% and 28%, respectively.

With CCS technology incorporated, the reduction in life cycle carbon emissions of one kWh of electricity reaches 82%, considering a 100-year time horizon. As previously done for ammonia production, the carbon footprint of 1 kWh electricity-based ammonia via SMR was evaluated considering a 20-year time frame. The results indicate a GHG emission reduction of 57%, from 2.2 kg CO\(_2\)-eq/kWh without CCS to 0.94 kg CO\(_2\)-eq/kWh with CCS. As denoted by Howarth and Jacobson [57], in terms of carbon footprint, the benefit of CCS is counterbalanced by methane emissions (direct and upstream) and their higher warming potential in the atmosphere over an integrated 20-year period.

For the impact category fossil depletion, a similar trend is observed, except for the scenario with CCS, where there is a slight increase. The extra energy needed to compress and inject the CO\(_2\) was modeled from the UK’s grid mix, which has more than 22% renewables and 23% nuclear [34]. Taking the UK country mix as a benchmark (0.145 kg oil-eq/kWh), using methane-based ammonia entails a rise of 6.5 to 7.4 times for this impact (without considering co-generation); in the other scenarios, the comparison ranges from 0.4 to 1.3.

Freshwater eutrophication potential is associated with the emissions of nitrogen (N), phosphorus (P), and other nutrient-rich substances that result in over-fertilization of freshwater bodies, such as lakes, streams, rivers, and reservoirs, among others. In the case of scenarios where renewable energy (wind, PV, and geothermal) is used to produce ammonia, the FEP is 3 to 6.6 times higher than the UK’s grid mix (0.058 g P-eq/kWh [34]). Emissions of phosphates during waste treatment processes of copper, hard coal, and
lignite mining are the most significant contributors. Considering the same benchmark, hydro-based ammonia, nuclear-based ammonia, and methane-based ammonia show better results (between 0.3 and 0.9 of the UK’s grid mix results). The use of surplus heat for district/industrial heating denotes marginal benefit.
Regarding the impact category ozone depletion potential, a similar tendency as for climate change and fossil depletion is observed, where methane-based ammonia entails a higher impact than electrolysis-based ammonia. The ODP is linked with stratospheric ozone destruction substances emissions, such as chlorodifluoromethane (HCFC-22) and bromochlorodifluoromethane (Halon 1211), present in the natural gas extraction and transport processes. For this reason, the positive effect of co-generation is more significant (i.e., with co-generation, there is a reduction in the ODP from 56% to 70%). Without the co-generation credits and considering the electricity production from a conventional natural gas power plant in the UK as the reference point (0.036 mg CFC-eq/kWh), the scenarios SMR-CCS-M-UK, SMR-M-UK, SMR-S-UK, and E-PV-M-CL have a higher impact, of 4.1, 3.8, 3.5, and 2.2 times, respectively. In the case of PV-based ammonia, the emissions of CFCs and HCFCs during the manufacturing of the PV cell components contribute to this impact.

Photochemical oxidant formation potential is related to the emission of nitrogen oxides and reactive hydrocarbons that form tropospheric ozone in the presence of sunlight, known as ‘smog.’ Nitrogen compound emissions (NO\textsubscript{X}, NO) are inherent in ammonia-based electricity production; in this study, the direct emissions during CHP plant operation contribute $9.6 \times 10^{-4}$ kg NMVOC-eq per kWh, which represents between 20% and 40% of the total impact for the studied scenarios. It is worth noting that the CHP cycle on which the model was based does not have NO\textsubscript{X} control equipment; therefore, a reduction of this impact would be expected for commercial operation. Emissions of internal combustion engines for land and sea transport also contribute to this impact, ranging from 4% to 48% of the total.

Terrestrial acidification is the change of soil chemical properties due to the accumulation of acidifying forms of nitrogen and sulfur; the air pollutants NO\textsubscript{X}, NH\textsubscript{3}, and SO\textsubscript{2} are the main precursors. Small-scale production is favored since there are no fugitive emissions of ammonia related to storage and transport. Ammonia emissions during transport represent 27% of the total impact for the closest location to the UK, Morocco, and up to 73% for Australia.

The excess energy emitted during the spontaneous disintegration of atoms (radioactivity) is known as ionizing radiation. Ionizing radiation exposure could come from natural and human-made sources, and the effect on human health (tissue or organ damage) de-
pends on the type, and dose of radiation received [60]. Nuclear-based ammonia entails the higher impact for this category, 19.8 times the UK’s grid mix (0.278 kg U235-eq/kWh electricity [34]), followed by methane-based ammonia at small-scale production, due to the inflow of electricity from the UK country mix (23% of nuclear [34]).

LCA studies on ammonia as a fuel for transport [11, 61–63] and power generation [12, 15, 16, 64] are less numerous than studies on ammonia production solely. Even though a straightforward comparison cannot be made due to differences in functional units, impact assessment methods, and in some cases, blends with other fuels, our results show a good agreement with those found in the literature. Bicer and Dincer [12] evaluated the environmental impact of ammonia from wind electrolysis for power generation. They found a reduction in GWP and ODP of 36% and 54%, respectively, while TAP increased 3969% relative to 1 kWh of electricity from a conventional natural gas power plant. In this study, without co-generation, the average impact of 1 kWh of ammonia-based on wind electrolysis is −41%, −36%, and 2629% relative to a natural gas conventional power plant, for GWP, ODP, and TAP, respectively. Razon and Valera-Medina [16] used pure methane as the benchmark for their LCA results of ammonia/methane blends in a gas turbine. A summary of the average percentage difference relative to benchmark they reported for GWP and those found in this study (without co-generation) are shown in Figure 7. The variability between the results of this study and those found in the literature [12, 16] corresponds to the uncertainty and variability inherent in all stages of LCA studies regarding system boundaries, assumptions, data sources, and impact assessment methods. However, an agreement is observed in the comparison with the literature, in sign and order of magnitude.

The efficiency of electrolysis-based ammonia is expected to increase in the medium-term future; that should come from the electrolysis step for hydrogen production [19]. More efficient scenarios were modeled for the UK and Morocco (Figure 8). These scenarios show a reduction (8–44%) in all impact categories for 1 kWh of ammonia-based electricity for the UK and Morocco. Current data show an average of 9.5 MWh is used to produce 1 t of liquid ammonia, whereas in the more efficient scenario, an energy input of 6.5 MWh per t of liquid NH₃ was included [19].

To contribute to the decarbonization of power and transport sectors, it is evident that ammonia via electrolysis is more favorable than ammonia via SMR and even with CCS. Considering the future demand for ammonia will likely be significantly higher by 2050, much more than the present supply capacity [39], a secondary analysis of the GWP of electrolysis-based ammonia using the grid mix and marginal electricity was developed. As seen in Figure 9, electricity production from electrolysis-based ammonia from non-
renewable sources entails a higher carbon footprint for all the studied scenarios. With a lower carbon grid mix, such as the UK’s, the GWP is lower, 1.835 kg CO$_2$-eq per kWh, but still much higher than that of the UK’s grid mix or a conventional natural gas power plant, 0.325 and 0.511 kg CO$_2$-eq per kWh, respectively. Considering a future more efficient electrolysis- and Haber–Bosch-based process, an average reduction of the GWP is projected (37%). However, the carbon footprint in the base case scenario is twice the conventional natural gas plant reference point. In comparison, with electricity from renewable and nuclear-based ammonia, the GWP is much lower.

![Figure 8](image)

**Figure 8.** Environmental profile of electricity production based on green ammonia. For the production of green ammonia, current data consider an energy input of 9.5 MWh per t NH$_3$; for the future 6.5 MWh per t NH$_3$ is considered.

![Figure 9](image)

**Figure 9.** Global warming potential results of co-generation cycle based on ammonia, FU: 1 kWh of electricity. (a) comparison of different locations’ grid mixes and marginal electricity (natural gas in gas turbine power plant); (b) comparison of current and future efficiency of the electrolyzer and Haber–Bosch processes. E: electrolysis; mix: grid mix, NG: marginal electricity—natural gas in a gas turbine, M: mid-scale 100,000 ton per annum ammonia plant, UK: United Kingdom, MA: Morocco, AU: Australia, CL: Chile.

### 3.3. Limitations and Recommendations for Further Research

As seen from the sensitivity analysis of GWP20, upstream processes and their associated emissions, such as the fugitive emissions of the extraction and distribution of natural gas, have an important contribution to the environmental profile of energy systems.
Therefore, using secondary information for these processes could imply a limitation of this study regarding the ammonia systems based on natural gas. In the literature, various researchers [65,66] observed a high variability (several orders of magnitude) of fugitive methane emissions and their impact on climate change associated with the upstream processes of the natural gas supply chain. The results of the natural gas-based ammonia and natural gas ammonia-based electricity could be higher if the fugitive emissions are significantly higher.

Even though the use of heat output for industrial or domestic applications enhances the studied cycle’s environmental profile, particularly for GWP, FDP, and ODP, further research is needed for systems that only produce electricity. This would deepen the understanding of the ecological implications of the use of ammonia for large power generation.

The characterization of the life cycle impacts of the power cycle was based on simulated results; therefore, as experimental or small-scale ammonia to power systems go into operation, real emissions and performance data should be used to update the present findings.

One of the by-products that could be obtained is water with high nitrogen levels, which could be used as fertilizer, and the potential benefit should be studied in more detail.

In the present study, transport and distribution were included using a LCI database with conventional technologies (fossil-based); however, transport’s contribution could be significant for specific impact categories (GWP, FDP, ODP, POFP), especially for renewable or nuclear-based ammonia, because of their low environmental impact. Further research is needed to integrate ammonia-based sea transport in the assessment. Indeed, as ammonia could become an essential factor in the energy mix, and electricity is a fundamental input in the life cycle of most of products, its impact should be reassessed, especially for very low-emission ammonia systems such as renewable or nuclear-based.

4. Conclusions

The environmental performance of a novel application of ammonia (NH₃) as an energy carrier was evaluated using a life cycle approach, comparing methane-based and water-based ammonia through various scenarios (combining different technologies, locations, and resources). This study shows that the ammonia production process is critical to any system that uses it. Two levels of analysis are presented: ammonia production and ammonia-based electricity production.

The results show that contributing significantly to the power sector’s decarbonization is only possible with electrolysis-based ammonia from renewable or nuclear sources. However, climate change is not the only thing that affects the environment, and if other environmental impacts are considered, there is a negative impact due to the use of ammonia from renewable or nuclear sources. This is due, in part, to the inherent emissions of nitrogen substances and the amount of energy needed for ammonia production. There could be a significant potential reduction of this negative impact with an increase in the efficiency of the electrolysis—Haber–Bosch technology. Nuclear-based ammonia performs very unfavorably regarding ionizing radiation potential.

CCS technology is seen as a viable option during the transition to a low-carbon energy market; however, the potential benefit could be less than previously estimated if a shorter time frame for GWP is used. This is especially significant in methane-based systems, such as in the production of blue ammonia and its applications, due to the higher contribution to global warming of methane emissions. Moreover, due to the extra energy needed for the process, blue ammonia’s environmental profile worsens compared with gray ammonia’s.

Although the results of this study are consistent with the rest of the literature, there may be some variability, due to the use of secondary information for upstream processes and simulated data for the power cycle, which may affect the impact indicators. Thus, further research is required on those critical aspects in order to fully grasp the benefits and drawbacks of ammonia as energy carrier.
In general, ammonia-based electricity production presents a favorable environmental profile, specifically with technologies such as hydropower and geothermal. However, this benefit could be improved by mitigating the emissions associated with the transport of ammonia. Consequently, the development of an ammonia production facility relatively close to the final users, a regional-level system, would enhance the positive effects of ammonia as an energy carrier. In addition, research and development is needed in order to analyze from a technical and environmental perspective the use of ammonia as a fuel for sea transportation, which would enhance the environmental performance of green ammonia.

Supplementary Materials: The following are available online at https://www.mdpi.com/article/10.3390/en14206721/s1, Table S1: Life cycle inventory (LCI) of ammonia production, FU 1 t of liquid NH$_3$, Table S2: Life cycle inventory of ammonia storage and transport, FU 1 t of liquid NH$_3$, and Table S3: Life cycle inventory of ammonia-based power generation, FU 1 kWh of electricity.

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Abbreviations

| Abbreviation | Description |
|--------------|-------------|
| ASU          | air separation unit |
| CCS          | carbon capture and storage |
| CFC          | chlorofluorocarbon |
| CHP          | combined heat and power |
| CO$_2$       | carbon dioxide |
| FDP          | fossil depletion potential |
| FEP          | freshwater eutrophication potential |
| GHG          | greenhouse gas |
| GWP          | global warming potential |
| H$_2$        | hydrogen |
| IRP          | ionizing radiation potential |
| ISO          | International Organization for Standardization |
| LCA          | life cycle assessment |
| LCI          | life cycle inventory |
| LCIA         | life cycle impact assessment |
| N            | nitrogen |
| NH$_3$       | ammonia |
| NMVOC        | non-methane volatile organic compounds |
| NO           | nitrogen monoxide |
| NO$_x$       | nitrogen oxides |
| N$_2$O       | nitrous oxide |
| ODP          | ozone depletion potential |
| P            | phosphorus |
| POFP         | photochemical oxidant formation potential |
| PV           | photovoltaic |
| SMR          | steam methane reforming |
| SO$_2$       | sulphur dioxide |
| TAP          | terrestrial acidification potential |
| UK           | United Kingdom |
| U235         | uranium-235 |
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