Quantifying crop pollinator dependence and its heterogeneity using multi-level meta-analysis

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Abstract
1. Biotic pollination can benefit crop production, but its effects are highly variable. To maximise benefits from this ecosystem service, we need a greater understanding of the factors that cause variation so that ecological intensification can be more effectively applied.
2. We focus on understanding the benefits of pollination to faba bean Vicia faba. We use a literature review followed by multi-level meta-analysis to estimate overall benefits of pollination to faba bean yield and to quantify variation (heterogeneity) in these benefits associated with different contextual factors (e.g. plant genotype, growing environment).
3. Our overall estimate of pollination benefit to faba bean yield, expressed as the percentage yield reduction without pollination, is 32.9% (confidence interval: 21%–43%). Based on the prediction intervals, which include the heterogeneity in pollination benefit, there is an 80% chance that pollination will increase yield of a faba bean crop.
4. Half of all heterogeneity in pollination dependence was due to differences between plant genotypes. The number of beans per plant showed similar pollination dependence to yield mass per plant while pod number and number of beans per pod underestimated yield benefits. There was weak evidence to suggest pollination benefits vary between pollinator species, with honeybees showing a smaller yield increase.
5. Differences in the experimental method used to assess pollination benefit did not significantly affect the estimate, including the growing environment, measurement scale, or whether the effects of experimental pollinator enclosures were controlled. This suggests that simplified experimental studies comparing yield of open-pollinated and enclosed plants can provide reliable insights into pollination benefits across a large range of plant genotypes and landscapes.
6. Synthesis and application. We found high variability in pollination benefits both between and within publications in our meta-analysis. Plant genotype, how yield was measured, and pollinator species affected the level of pollination benefit. Despite variability in pollination benefits due to various contextual factors (both inside and outside of grower control), there is a high likelihood that biotic pollination...
INTRODUCTION

The contribution of biotic pollination to crop yield has been recently valued at $235–577 billion globally each year (Potts et al., 2016). Understanding this ecosystem service can help us to maximise those pollination benefits by guiding effective decision-making in policy, crop breeding and agronomy. This ecological intensification offers a more sustainable route to increased crop productivity than the use of conventional inputs and agrochemicals (Bommarco et al., 2013). However, there is significant variation in the contribution of pollinators to crop yield; it varies around the world due to differences in the crop types cultivated (Potts et al., 2016), the diversity and abundance of pollinator communities (Woodcock et al., 2019), and many other contextual factors (Garibaldi et al., 2018; Tamburini et al., 2019). It is essential to quantify and explain this variation so that ecological intensification can be more optimally applied in specific contexts. With a better understanding, pollination could increasingly be considered an agronomic input that can complement or replace conventional inputs (Garratt et al., 2018; Tamburini et al., 2017) with more predictable returns for investments in on-farm ecological management (Kleijn et al., 2019). We advocate and demonstrate the use of multi-level meta-analysis to understand variability in pollination benefit and predict the range of pollination dependence across different contexts.

Isolated experimental and observational studies are inadequate to assess and understand the benefits of biotic pollination to a crop. The primary benefit of biotic pollination to plants is via increased ovule fertilisation, and an increased number of seeds for a plant to develop. If a plant has sufficient resources, there can be a corresponding increase in crop yield (Garibaldi et al., 2018). High soil nutrient availability can therefore make plants more responsive to pollination (Garratt et al., 2018). Conversely, adverse weather or herbivory can suppress the benefits of pollination by damaging floral organs and developing seeds, or by causing resource limitation (Klein et al., 2015; Lundin et al., 2013). Each isolated study represents a specific combination of nutrient availability, weather and pest pressure, all of which limit the applicability of the findings to different contexts (Garibaldi et al., 2018; Tamburini et al., 2019). Quantitative research synthesis, using ‘review-generated’ evidence (Cooper, 2015), is therefore needed to understand and account for the variation associated with these different contextual factors.

To encourage ecological intensification by farmers and reap its associated environmental benefits, it is important to report crop responses in economically relevant terms (Kleijn et al., 2019). The use of different metrics to measure crop yield responses to pollination across the literature therefore presents a challenge. Some studies may be concerned only with fertilisation and therefore present abstract fertilisation metrics such as the number of seeds or pods per flower (Drayner, 1959). Due to time or labour constraints, other studies may forgo yield mass measurement and only record simpler metrics such as the number of seeds in a representative sample of pods. The relationships between different yield components can be nonlinear (Bishop et al., 2020) and as discussed above, contextual factors can modify the translation of pollination into crop yield and economic output. A better understanding of the relationships between different yield metrics could help practitioners to generalise across findings and apply literature from more relevant contexts.

To determine the effect of biotic pollination, it is typically necessary to enclose plants or inflorescences to stop pollinators visiting the flowers. These physical enclosures could affect estimated pollination benefit by preventing pest access, increasing local air temperature and humidity, or reducing light available to the plant. Alternative experimental designs that control for the effect of enclosures may involve growing all plants in larger enclosures and introducing managed pollinators or conducting hand pollination treatments (Bishop et al., 2020; Poulsen, 1977). This is unlikely to represent the activity of local pollinator communities and the yield benefits encountered in the field (Woodcock et al., 2019). Different experimental designs therefore represent different compromises between on-farm relevance and the level of experimental control. These compromises could have far-reaching implications for experimental findings and their application (e.g. see Ainsworth et al., 2008). A greater understanding of how methodological differences translate into pollination dependence estimates could help optimise experimental design and between-study comparisons.

Different crop species vary in their dependence on biotic pollination; some species produce no seeds without external pollination while others show more nuanced responses (Aizen et al., 2009). Different plant genotypes within crop species also vary in their response to additional pollination (Marini et al., 2015) though this has rarely been considered in economic analyses (but see Fijen et al., 2018 and Klatt et al., 2014). Different plant genotypes are cultivated in different regions, due to a range of factors including consumer preferences and climate requirements. Understanding the range of pollination dependence within a crop (including in breeding material) is necessary to produce a robust estimate of the current and potential future extent of pollination dependence.
In this paper, we focus on faba bean \textit{Vicia faba} L., a partially pollinator-dependent grain legume that has a high protein content and is used as a human food and in livestock feed. Faba bean itself has a valuable role in ecological intensification, supporting the growth of other crops by increasing soil nutrient availability, and supporting pollinating insects by provision of floral resources (Köpke & Nemeczek, 2010). The crop produces seed by three mechanisms; autonomous self-fertilisation (autofertility), self-fertilisation mediated by insect visitation and cross-fertilisation following pollen transfer by insects (Drayner, 1959). The relative contribution of each mechanism is known to vary with genotype and the level of cross-fertilisation in previous generations (Frusciante & Monti, 1980). Various authors over the last 60 years have recommended that the crop be supplemented with insect pollinators using domesticated pollinators or pollinator habitat creation (Cunningham & Le Feuvre, 2013; Kendall & Smith, 1975; Riedel & Wort, 1960) but this has not become commonplace.

The variation in pollination benefits to yield, in faba bean and other crops, makes it difficult for practitioners to understand the likely benefits of management for pollinators on a given farm and presents a barrier to ecological intensification and its associated benefits (Kleijn et al., 2019). Meta-analysis is a commonly used approach to synthesize across different sources of evidence and provide an overall effect size estimate (Higgins et al., 2019). However, rather than investigating or accounting for complexity, meta-analyses often simplify data, analysing only a limited subset of available data (or averaging across it) to reduce problems associated with dependence (e.g. repeated measurements across years or different yield metrics; Noble et al., 2017). Recent advancements in meta-analytic approaches now allow us to include more data while accounting for its dependence.

We use such an approach, first using a literature review to identify all experimental estimates of faba bean pollination dependence, followed by application of multi-level meta-analytic models (e.g. Nakagawa & Santos, 2012). We quantify the effect of different factors such as pollinator species on the benefit of pollination via fixed effects. Our approach also allows us to quantify the variation (heterogeneity) associated with other contextual factors via random effects (Nakagawa & Santos, 2012). Furthermore, rather than just simplifying heterogeneity to statistics, we demonstrate that using prediction intervals (IntHout et al., 2016) we can use our understanding of heterogeneity to predict the likely range of pollination dependence across other contexts.

2 | MATERIALS AND METHODS

2.1 | Identifying publications

To identify relevant publications for our meta-analysis, we searched Web of Science and Scopus using the terms ‘faba’ and ‘pollination’ in November 2018. To include grey literature, we conducted a search of Google Scholar using the terms ‘faba’ and ‘pollination’ in December 2018; we sorted results by relevance, then checked the first 200 articles (Haddaway et al., 2015). For robustness, we then screened the reference lists of publications that had met our inclusion criteria. We only included publications in our meta-analysis if they compared yield production between plants receiving a pollination treatment and plants that did not receive a pollination treatment. We only included publications that presented yield using at least one of four metrics most likely to correspond to economically relevant yield—yield mass, bean number, pod number or bean number per pod (Bishop et al., 2020). Several publications reported yield using other metrics that do not translate to economically relevant yield production, these were not included in the meta-analysis but are summarised in Table S1. We included publications that represented insect pollination by hand pollination defined as either tripping (pulling apart keel and wing petals) or hand pollination (tripping + transfer of external pollen). One person (JB) conducted the literature review, we include further details about the review and a PRISMA diagram in the Supporting Information.

2.2 | Calculating effect sizes

To quantify the effect of biotic pollination on faba bean yield (the effect size), we used the natural log of the response ratio (lnRR), which is the log proportional change in yield between plants receiving a pollination treatment and undisturbed plants (Hedges et al., 2016).

This converts to the proportion of yield lost without pollination using the simple formula $1 - \exp(\lnRR)$.

2.3 | Weighting effect sizes

A meta-analysis weights effect sizes by the inverse of their sampling variance. Where standard deviation (or a convertible alternative) was missing from publications, we requested this data from authors for papers dated 2000 onwards. When a measure of variance was not obtainable, we imputed standard deviation based on the fitted relationship between mean, SD and $n$ in the available data (91 data rows imputed; adjusted-$R^2$ of these models were between 0.29 and 0.46). If a variance measure was not available and we could not impute it, then the effect size was excluded from our meta-analysis (Table 1; Table S1). See Supporting Information for more details and a sensitivity analysis ran without imputed SD; the results were similar.

In medical and eco-evolutionary meta-analyses, $n$ is typically equal to the number of individuals tested in a given experiment. For many effect sizes in our analysis, there was a difference between the number of experimental units (e.g. field patches, or cages containing plants) and the number of individual plants. Where available, we collected information about both sample size types. We conducted our primary analysis using $n$ as the number of experimental units and include findings for $n$ as number of individuals in the Supporting Information; the results were similar.
| Publication | Country | Source(s) of variation | Pollinator | Yield mass | Bean number | Pod number | Beans per pod |
|-------------|---------|-----------------------|------------|------------|-------------|------------|---------------|
| Riedel and Wort (1960) | UK | 2 years | Apis; Open | | 34 (24–54) || |
| Scriven et al. (1961)*\† | UK | 3 years | Apis; Open | 48 (39–62) | | | |
| Free (1966)\† | UK | 2 genotypes | Apis; Open | 46 (14–71) | 44 (20–63) | 22 (−2 to 43) | 13 (6–19) |
| Hanna and Lawes (1967)\† | UK | 2 years; 15 genotypes | Open; Trip | 44 (−122 to 91) | | | |
| Kendall and Smith (1975)\† | UK | 2 years; 15 genotypes | Bomb.; Apis; Trip | | | | |
| Poulsen (1977)\† | UK | Trip | | | 67 | 23 | |
| Frusciante and Monti (1980)*\† | Italy | 2 genotypes | Apis; Open | 51 (37–64) | | | |
| Stoddard (1986)* | Australia | Open | | 48 (37–60) | 44 (30–57) | 8 (5–10) | |
| Dekhuizen et al. (1988) | Netherlands | Hand | −4 | 6 | −33 | | |
| Varis and Brax (1990)\† | Finland | 2 years | Apis; Open | 60 (13–88) | 61 (44–74) | 49 (19–66) | 20 (10–32) |
| Mesquida et al. (1990) | France | 2 genotypes | Bomb.; Apis; Hand; Open | 32 (−26 to 83) | 38 (−24 to 89) | 14 (−4 to 28) | |
| Le Guen et al. (1993)\† | France | 2 genotypes | Bomb.; Trip; Open | 36 (−24 to 74) | 42 (−27 to 82) | | |
| Kolowski (1996)*\† | Poland | 3 years, 7 genotypes | Open | 19 (−23 to 49) | 11 (−1 to 19) | | |
| Somerville (1999) | Australia | Apis; Open | | 22 (12–29) | | | |
| Ghamdi and Ghamdi (2003)\† | Saudi Arabia | Apis; Open | | 27 (21–34) | 22 (17–28) | | |
| Musallam et al. (2004)*\† | Jordan | 4 genotypes | Open | 33 (32–34) | 6 (4–9) | 13 (12–14) | |
| Benachour et al. (2007) | Algeria | Open | | 67 | 50 | 25 | 10 |
| Aouar-sadii et al. (2008) | Algeria | Open | | 16 | 13 | −47 | 41 |
| Garratt et al. (2014) | UK | Bomb., Hand, Apis | | | | | 12 (7–20) |
| Suso and del Rio (2015)*\† | Spain | 6 genotypes | Open | 5 (−48 to 37) | | | |
| Birkin and Goulson (2015)*\† | UK | Open | | 69 | 50 | 76 | |
| Nayak et al. (2015) | UK | Open | | | | | 40 |
| St-Martin and Bommarco (2016) | Sweden | 4 soil types | Bomb. | 30 (12–52) | | | |
| Bishop et al. (2016) | UK | 3 stress; 5 stress | Bomb. | 17 (7–38) | 15 (−1 to 29) | 3 (−11 to 16) | 13 (0–22) |
| Bishop et al. (2017) | UK | 2 stress; 2 environ. | Bomb.; Trip; Open | 39 (5–61) | 43 (6–65) | | |
| Kyllönen (2018) | Finland | Apis; Open | | | 39 (17–62) | | |
| Bishop et al. (2020) | UK | 2 years; 5 genotypes; 2 environ. | Hand; Trip; Open | 6 (−167 to 68) | 12 (−153 to 82) | 1 (−160 to 64) | |
2.4 Multi-level meta-analysis models

The publications identified in our literature review were diverse; for example, they were conducted in different countries, they used different pollinator species and different cultivars (plant genotypes), and they measured yield in different ways. The publications often reported more than one effect size; for example, they compared several cultivars, repeated experiments across several years or reported yield using several metrics. The 22 publications that satisfied our inclusion criteria for the meta-analysis (Table 1) included a total of 277 effect sizes. These differences within- and between-publications allowed us to investigate what causes variation (heterogeneity) in pollination benefit.

Including more than one effect size from the same publication (or from publications that are related in some way, e.g. because they tested the same cultivar) can however lead to correlations (clustering) among these effect sizes, which invalidates model assumptions of independence (Noble et al., 2017). Throughout, we used multi-level meta-analytic models to account for this dependence via random effects and with sampling variance–covariance matrices. These multi-level models follow the same principles as linear mixed effects models (LMMs; Harrison et al., 2018).

To estimate the overall pollination dependence of faba bean, we use a multi-level meta-analytic model with only random effects (a null model, analogous to an LMM with no fixed effects). We identified the optimal random effects structure by comparing AIC of different candidate models. The random effects we tested were an individual effect size identifier (unique per data row, necessary to estimate residual heterogeneity), a publication identifier, year nested within publication (where a publication reported experiments conducted across multiple years), cultivar (plant genotype; a partially crossed-random effect as five cultivars were tested across different publications), the country in which the experimental work took place and the author team (publications sharing an author may have methodological similarities that we did not capture with the other random effects). Apart from the effect size identifier, all the random effects were potential clustering factors. The optimal model is model RE0 in Table 2. We estimated total heterogeneity in pollination benefit and that associated with different levels of clustering (random effects) using $I^2$ (Higgins & Thompson, 2002).

As well as letting the model estimate dependence of effect sizes due to clustering (random effects, see above), we also explicitly modelled dependence of sampling error variance (Noble et al., 2017) because random effects do not control for dependence arising from sampling errors that are shared among effect sizes (i.e. sampling error co-variances). In some experiments, different treatments were compared to a shared control treatment (most commonly,

| Model | Moderators (fixed effects) | Random effects |
|-------|--------------------------|----------------|
| RE0   | NA                       | Publication, year within publication, cultivar, individual effect size (residual) |
| RE1   | NA                       | Publication, cultivar, individual effect size (residual) |
| RE2   | NA                       | Country, publication, year within publication, cultivar, residual |
| RE3   | NA                       | Authors, publication, year within publication, cultivar, residual |
| RV4   | Yield metric type with 4 levels (pod number, yield mass, bean number, bean per pod) | Publication, year within publication, cultivar, individual effect size (residual) |
| RV3   | Yield metric type with 3 levels [pod number, (yield mass & bean number), beans per pod] | |
| Po5   | Pollinator type with 5 levels (honeybees, bumblebees, hand poll, tripping, open) | |
| Po2   | Pollinator type with 2 levels [honeybees, (bumblebees & hand poll & tripping & open)] | |
| Ba3   | Bagging control with 3 levels (controlled, semi-controlled, not controlled) | |
| Fi2   | Plant growing conditions with 2 levels (field or plant pot) | |
| Cu2   | Type of cultivar with 2 levels (commercial or breeding) | |
| Sc3   | Scale of experimental aggregation with 3 levels (within plant, plant, plot) | |
| Ye0   | Publication date as a continuous variable | |

TABLE 2 Summary of all statistical models reported in the manuscript, presented in order of appearance. Likelihood ratio tests were used for all model comparisons. For associated R code, see Table S2.
different pollinator species compared to an excluded control group. We addressed this dependence by dividing sample size for the control group evenly among the shared comparisons before calculating the effect sizes (Higgins et al., 2019). In other cases, multiple effect size estimates were provided per individual (or experimental unit) because multiple yield metrics were measured. We used variance-covariance matrices to account for the resulting correlated sampling (error) variance (Noble et al., 2017). In the main text, we present results from a conservative model that assumed a correlation of 0.5 between effect size sample variances that were measured on the same experimental units. In the Supporting Information, we present results from more conservative models that assume a correlation of 0.9, the results were qualitatively similar.

We added moderators (fixed effects) to the optimal random effect model (RE0) to try to explain some of the heterogeneity in effect sizes and quantify how different factors affect the level of pollination benefit. These models are equivalent to LMMs with fixed and random effects. The moderators we tested were (a) response metric (e.g. yield mass, bean number), (b) pollinator type (e.g. bumblebee, honeybee), (c) whether effects of pollinator enclosures were controlled; (d) plant culture method (pot-grown or field-grown plants); (e) whether measurements were made at a within-plant scale, plant scale or field-area scale and (f) year of publication. We established the statistical significance of moderators by comparing models including each moderator to the null model (RE0) via likelihood ratio test. We ran separate models for each moderator and did not test simultaneously including each moderator to the null model (RE0) via likelihood while results are reported from models fitted with restricted maximum likelihood. All models that we report results from are documented in Table 2, which corresponds to Table S2 in the Supporting Information where R code is provided for each model.

We visualised results from our models using orchard plots which show 95% confidence intervals, CIs, and 95% prediction intervals, PIs (Nakagawa et al., 2020). Confidence intervals relate to the most likely location of the cross-study average effect. In contrast, prediction intervals incorporate heterogeneity and relate to the expected range of pollination dependence likely to be found if an additional experiment were conducted or if a farmer grew a faba bean crop (IntHout et al., 2016).

2.5 Publication bias and sensitivity analysis

Our meta-analysis uses existing literature and it is important to check for biases in that literature (e.g. only publishing significant results). We tested for bias in two ways: (a) visual inspections for a pattern of asymmetry in a funnel plot of the ‘meta-analytic residuals’ from the meta-regression model including all significant moderators (Nakagawa & Santos, 2012) and (b) testing funnel asymmetry using a multi-level model version of Egger’s regression by including sampling standard error, SE, as a moderator in the null models and a full model containing all significant moderators (Egger et al., 1997; Moreno et al., 2009); if sampling SE is significant, it statistically supports asymmetry in the funnel. In a funnel plot, effect sizes are plotted against their associated precisions; in the absence of bias the effect sizes with lower precision would typically, by chance, scatter more widely, forming a symmetrical inverted funnel centred on the mean effect, while if the effect sizes are biased, the funnel would be skewed or asymmetrical (see Egger et al., 1997). These two methods, which we used, deal with non-independency in our dataset, while the original funnel plots and Egger’s regression does not (Nakagawa & Santos, 2012). Our moderator test for year of publication can also be used to look for time lag bias (or decline effect, Koricheva & Kulinskaya, 2019); that is, a significant year effect could indicate a systematic change in effect size over time.

3 RESULTS

Table 1 summarises all publications included in our meta-analysis. Several other publications only partially met our inclusion criteria so were not included in the meta-analysis, these are summarised in Table 1 (marked with asterisks) or Table S1.

Faba bean plants not receiving a pollination treatment produced on average 32.9% (CI 21%–43%) less yield than pollinated plants (RE0 in Table 2; Figure 1). If we assume that faba bean crops are optimally pollinated, this indicates that 1.5 million tonnes of faba bean production globally is due to insect pollination (Table 3).

There was very high heterogeneity in pollination dependence both within and between publications (from RE0; \( R^2_{total \, null \, model} = 0.918, \ R^2_{studystudy} = 0.054, \ R^2_{yearinstudy} = 0.074, \ R^2_{cultivar} = 0.484, \ R^2_{residual} (residual) = 0.305 \)). The prediction interval is a highly skewed distribution when transformed to the scale of percentage yield loss without pollination (PI \( = -69.5\% \) to 73.5%; Figure 1) so we consider it more useful to calculate the probability that a farmer would experience a benefit of biotic pollination to faba bean yield (area under the curve beyond zero; Figure 1), which is 80%.

Plant genotype (cultivar) accounted for more than half of the heterogeneity between effect sizes. If heterogeneity due to cultivars is removed (see Supporting Information), the prediction interval for pollination dependence becomes substantially smaller (−28% to 65%) translating to an 89% chance of seeing a benefit (Figure 1). Responses to pollination were different when experiments were repeated across more than 1 year (likelihood ratio test (LRT) comparing model RE0 and RE1, \( p = 0.041 \)); the heterogeneity associated with different years within publications was approximately double that between publications (see \( R^2 \) above). The country in which experiments were conducted (RE2), or the authors of a publication (RE3), did not explain a significant amount of variation (Supporting Information).

Pollination dependence varied significantly between different metrics for measuring yield (LRT RV4 vs. RE0, \( p < 0.001 \);
Bean number and yield mass had similar estimates (LRT RV4 vs. RV3, \( p = 0.091, R^2_{\text{marginal}} = 0.066 \)) with a combined estimate of 37% reduction without pollination (RV3; PI = −51.9% to 73.9%) while pod number and number of beans per pod were significantly less responsive to biotic pollination.

### FIGURE 1 Overall estimate for pollination dependence of faba bean.

Panels a and c show confidence intervals and prediction intervals, including an illustration of the prediction interval if heterogeneity caused by cultivars was not present (by estimating 95% prediction interval without the variance component due to cultivars) on the lnRR and response ratio (RR) scale, respectively. Panels b and d show orchard plots for the random effects model; the thin horizontal line shows the prediction interval, the bold horizontal line shows the confidence interval, and the point shows the mean estimate on the lnRR and RR scale.

### TABLE 3 Translating yield dependence into global estimates of pollination value.

We calculate this by multiplying the proportion pollination dependence by the global production (4.6 Mt annual mean global production 2009–2018), which comes from FAOSTAT accessed 28/02/2020.

| Model                        | Pollination dependence (confidence interval) | Pollinator-dependent yield (confidence interval) |
|------------------------------|----------------------------------------------|-----------------------------------------------|
| Average                      | 32.9% (21–43%)                               | 1.51 Mt (0.97–1.98 Mt)                         |
| Commercial cultivars only    | 31.3% (21–41%)                               | 1.44 Mt (0.97–1.89 Mt)                         |
| Yield mass and bean number only | 37% (27–46%)                             | 1.69 Mt (1.24–2.12 Mt)                         |
The scale of yield measurement did not significantly affect the estimate though there were few effect sizes available to make this comparison, with the majority of effect sizes measured at a whole-plant scale (LRT Sc3 vs. RE0, \( p = 0.725 \)).

There is weak evidence that pollination benefits vary with pollinator species. While four of the five pollinator groups tested had similar estimates, honeybees were less effective (LRT Po2 vs. RE0, \( p = 0.110, R^2_{marginal} = 0.008 \); Figure 3).

We found no difference between estimated pollination dependence when effects of enclosure were controlled (e.g. all plants inside an enclosure) or uncontrolled (e.g. enclosed plants vs. open plants; LRT Ba3 vs. RE0, \( p = 0.720 \)). Likewise, pollination dependence...
did not change if plants were grown in the field or a pot (Supporting Information).

Whether the plants tested were commercially available cultivars or breeding lines did not significantly change the dependence estimate (LRT Cu2 vs. RE0, \( p = 0.618 \)), but there was a greater range of effect sizes in the breeding lines (Figure 4; F test with null hypothesis that variances were equal, \( p < 0.001 \), ratio between variances 0.19).

There was little evidence of publication bias. In a model using only non-imputed SD values, and containing moderators of \( \sqrt{v_i} \); where \( v_i \) = sampling variance for effect size, so \( \sqrt{v_i} \) is standard error, year of publication, and yield metric (3 levels), the slopes of \( \sqrt{v_i} \) and year were not significant (\( p = 0.791 \) and \( p = 0.080 \) respectively; Supporting Information). Using the dataset that included imputed SD values, there was a significant trend of decreasing pollination dependence over time (LRT Ye0 vs. RE0, \( p = 0.036 \); Figure 5).

4 | DISCUSSION

Our results show that faba bean loses an average of 32.9% yield without biotic pollination. Multi-level meta-analysis allows us to quantify variability (heterogeneity) around the average estimate and attribute this to different sources (Nakagawa & Santos, 2012). The confidence interval, 21%–43%, shows the most likely location of the cross-study average effect and we suggest this information is used in large-scale economic valuations of pollination services (e.g. Table 3). We found high heterogeneity in pollination dependence. The prediction interval (−70% to 74% in our overall model) incorporates the heterogeneity between effect sizes to illustrate the pollination benefits that a future experiment (or farmer growing a crop of beans) may find (IntHout et al., 2016). Confidence and prediction intervals are distributions, and they become highly skewed on the response ratio (RR) scale (Figure 1). We can, more usefully, calculate that there is an 80% probability that a given farmer will see a benefit of biotic pollination to faba bean yield. Using multi-level meta-analysis and reporting of prediction intervals or their derivatives in this way gives an honest appraisal of biotic crop pollination benefits to farmers.

More than half of the variation in pollination dependence was due to plant genotype. Authors calculating the economic value of biotic pollination commonly account for differences between crop species (Aizen et al., 2009), but there is a clear need to consider differences within crop species (e.g. Bishop et al., 2020; Klatt et al., 2014) when...
placing economic values on pollination services and/or planning pol-
licies and agronomic management. Plant genotype is something that
growers can control by choosing to grow a certain cultivar. Several
publications in our meta-analysis set out to compare the pollina-
tion dependence of different genotypes. Variability in pollination
dependence was much greater between (pre-commercial) breed-
ing lines than commercially available cultivars (Figure 4; Supporting
Information). If we re-calculate the prediction intervals without the
heterogeneity associated with cultivars (e.g. we assume everyone
grows a cultivar with average dependence), pollination dependence
becomes more predictable (PI –28 to 65%) and the likelihood of a
pollination benefit to yield increases from 80% to 89%. Publications
have identified faba bean genotypes that are unaffected by polli-
nator exclusion, and authors have begun to identify heritable traits
associated with this autofertility (Torres et al., 1993). Crop breeders
might wish to produce faba bean cultivars that do not depend on bi-
otic pollination, as this would remove one possible constraint on pro-
duction (Marini et al., 2015) but this is difficult to achieve in practice
(see Supporting Information). More experimental research is needed
to understand how and why pollination dependence varies between
plant genotypes and how this interacts with other factors such as
maximal yield potential and disease resistance. There has been a
reduction in pollination dependence over the 60-year period con-
sidered in our analysis, but it is not possible to distinguish between
biological changes in dependence and variations in experimental de-
sign or publication bias (Koricheva & Kulinskaya, 2019).

The prediction interval overlaps with zero, suggesting that some-
times, biotic pollination will not benefit faba bean yield.
Multiple factors can limit yield simultaneously (see Garibaldi et al.,
2018) and experiments, or growers, may find no benefit of pollin-
ation because various other contextual factors are limiting yield.
If a plant does not have sufficient resources to mature seeds, then
increases in ovule fertilisation may not translate into increases in
yield (Garratt et al., 2018). Likewise, if plants are damaged by pests
or adverse weather, then this could nullify any benefit of increased
ovule fertilisation, or even reduce efficiency if the plant has already
invested in a larger number of flowers and seeds (Melathopoulos
et al., 2014). Publications repeated across different years specif-
cally discussed variation in weather between years (e.g. Varis &
Brax, 1990). Variation in dependence was also reported between
soil types (St-Martin & Bommarco, 2016) and following heat stress
(Bishop et al., 2016). We found high heterogeneity between publica-
tions, between different years within publications, and high (re-
sidual) heterogeneity that we could not explain. This finding means
that the average level of pollination benefit will be specific to a
grower’s context (e.g. their soil type, akin to between publication vari-
ation) but other factors will cause additional variation within
that context (e.g. differences in weather). Some publications re-
ported negative impacts of pollination, but this was mostly where
there was no significant difference between pollination treat-
ments (e.g. Dekhuizen et al., 1988, but see Bishop et al., 2020).
These findings suggest that new work to quantify crop pollination
dependence (e.g. in other crops or cultivars where meta-analysis
is not currently possible due to limited prior research) needs to be
repeated across multiple years and/or sites to determine the range
of pollination benefit likely to be experienced.

We found weak evidence that the benefit of pollination is greater
when the pollinator is not honeybees. This supports ecological in-
tensification, for example, through habitat improvement for wild
pollinators, compared to a more conventional input-based approach
of introducing honeybee hives on to farmland. Several publications
in our analysis directly compared honeybees with other species and
found them to be less effective, requiring repeated floral visits to
achieve the same podset as some bumblebees (Garratt et al., 2014;
Kendall & Smith, 1975, but see Cunningham & Le Feuvre, 2013).
Several authors also reported differences in efficacy between bun-
blebee species, though we could not test this in our analysis (Kendall
& Smith, 1975; Le Guen et al., 1993). Differences in pollinator ef-
cicacy likely relate to differences in foraging behaviour, the vol-
ume or location of pollen held on the insect, their between-flower
movement and the frequency of flower visits made (Marzinzig et al.,
2018). Agronomic interventions targeted at specific pollinator spe-
cies, for instance sowing field margins with plants of a specific floral
structure, may be an effective means to attract species of particular
benefit to the crop (Garibaldi et al., 2015).

Our results show that there is no yield penalty associated with
enclosing plants to exclude pollinators. It is possible that enclosures
exerted a combination of positive and negative effects (and this may
vary with context, e.g. level of pest pressure), resulting in no differ-
ence in overall estimate. We have been unable to investigate more
nuanced potential effects of methodologies, for example, non-con-
trolled enclosures were always used with open pollination and com-
mercial genotypes. Regardless, there was no significant difference
in pollination dependence between experiments that controlled for
bagging effects and those that did not. Given the importance of con-
text and plant genotype, and the clear need to repeat trials across
multiple years and/or sites, this finding could help studies to be more
efficient (though this should be confirmed in other crop species).
Highly controlled experiments are resource intensive and this lim-
its the number of plant genotypes or contexts that they can cover.
Such experiments may also use pollinators that do not reflect the
abundance and activity of insects on-farm. To estimate the bene-
fits of pollination on-farm and to gain useful agronomic insights (e.g.
where to establish new pollinator nesting habitat or floral resources),
a comparison between open-pollinated plants and enclosed plants
is most likely to represent the pollinator community present (e.g.
Woodcock et al., 2019). Large-scale, simplified experiments using
this design could provide further insights into the role of context on
crop pollination benefits. Growers themselves may even be encour-
gaged to conduct these types of experiments (Garratt et al., 2019)
which could increase uptake and suitability of ecological intensifica-
tion practices (Kleijn et al., 2019).

In summary, we have used a repeatable and rigorous literature re-
view to identify all available comparisons of faba bean yield with and
without biotic pollination, and we have used multi-level meta-anal-
ysis to quantify and predict faba bean pollination dependence.
Given the importance of context that is increasingly recognised (e.g. Tamburini et al., 2019), quantitative research synthesis is necessary in valuations of crop pollination services. We found high heterogeneity in pollination benefit due to various contextual factors that are both inside and outside of grower control. While the benefits that growers receive from biotic pollination vary across space and time, there is a high likelihood that biotic pollination will increase faba bean yield. This strengthens the case for management of pollinators to maximise those pollination benefits.

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AUTHORS’ CONTRIBUTIONS
J.B. conceived the study, conducted the systematic review and meta-analysis and wrote the first draft of the manuscript and Supporting Information documentation. S.N. advised on the analysis, produced the figures and contributed to manuscript.

DATA AVAILABILITY STATEMENT
Data available via the Centre for Open Science https://doi.org/10.17605/osf.io/b3e9u (Bishop & Nakagawa, 2020).

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