DNA metabarcoding reveals introduced species predominate in the diet of a threatened endemic omnivore, Telfair’s skink (Leiolopisma telfairii)

Maximillian P. T. G. Tercel\textsuperscript{1,2} | Rosemary J. Moorhouse-Gann\textsuperscript{1,2,3} | Jordan P. Cuff\textsuperscript{1,4} | Lorna E. Drake\textsuperscript{1} | Nik C. Cole\textsuperscript{2,5} | Martine Goder\textsuperscript{5} | Rouben Mootoocurpen\textsuperscript{5} | William O. C. Symondson\textsuperscript{1}

\textsuperscript{1}School of Biosciences, Cardiff University, Cardiff, UK
\textsuperscript{2}Durrell Wildlife Conservation Trust, Trinity, Jersey
\textsuperscript{3}Department of Animal & Plant Sciences, NERC Biomolecular Analysis Facility, Sheffield, UK
\textsuperscript{4}Rothamsted Insect Survey, Rothamsted Research, Harpenden, UK
\textsuperscript{5}Mauritian Wildlife Foundation, Vacoas, Mauritius

Correspondence
Maximillian P. T. G. Tercel, School of Biosciences, Cardiff University, Sir Martin Evans Building, Museum Avenue, Cardiff, CF10 3AX, UK.
Email: max.tercel@hotmail.com

Funding information
Natural Environment Research Council, Grant/Award Number: NBAF983, NE/K00719X/1 and NE/L002434/1; Durrell Wildlife Conservation Trust, Grant/ Award Number: MR/S502455/1; British Herpetological Society, Grant/Award Number: 517513

Abstract
Introduced species can exert disproportionately negative effects on island ecosystems, but their potential role as food for native consumers is poorly studied. Telfair’s skinks are endemic omnivores living on Round Island, Mauritius, a globally significant site of biodiversity conservation. We aimed to determine the dietary diversity and key trophic interactions of Telfair’s skinks, whether introduced species are frequently consumed, and if diet composition changes seasonally between male and female skinks.

We used DNA metabarcoding of skink fecal samples to identify animals (COI) and plants (ITS2) consumed by skinks. There were 389 dietary presence counts belonging to 77 dietary taxa found across the 73 Telfair’s skink fecal samples. Introduced taxa were cumulatively consumed more frequently than other categories, accounting for 49.4% of all detections, compared to cryptogenic (20.6%), native (20.6%), and endemic taxa (9.5%). The most frequently consumed introduced species was the ant, \textit{Pheidole megacephala}, present in 40% of samples. Blue latan palm, \textit{Latania loddigesii}, was the most frequently consumed endemic species, present in 33% of samples but was only detected in the dry season, when fruits are produced. We found a strong seasonal difference in diet composition explained by the presence of certain plant species solely or primarily in one season and a marked increase in the consumption of animal prey in the dry season. Male and female skinks consumed several taxa at different frequencies. These results present a valuable perspective on the role of introduced species in the trophic network of their invaded ecosystem. Both native and introduced species provide nutritional resources for skinks, and this may have management implications in the context of species conservation and island restoration.

Keywords
dietary analysis, invasive species, island restoration, multiple markers, \textit{Pheidole megacephala}, reptiles, Round Island Mauritius
1 | INTRODUCTION

A novel species introduced into a new ecosystem can interact with taxa already present in a variety of ecological roles, for example, as a mutualist (Kaiser-Bunbury et al., 2011), competitor (Cole & Harris, 2011), predator (O’Dowd et al., 2003), prey (Li et al., 2011), or parasite (Arbetman et al., 2013). Introduced species are often associated with network and community restructuring (Memmott et al., 2000; Russo et al., 2014), as well as native biodiversity declines (Clavero & García-Berthou, 2005; Luque et al., 2014) and even ecological collapse (O’Dowd et al., 2003). However, introduced species may have more nuanced effects on ecosystems, including interactions beneficial to native species (Schlaepfer et al., 2011). For example, non-native trees may provide nesting sites to threatened birds (Schlaepfer et al., 2011) and non-native plants may provide floral resources to a range of threatened native pollinators (Baldock et al., 2015).

Introduced species have been studied extensively as invasive predators and herbivores, but their role in the diet of native species has been given less attention. A few studies examine this subject explicitly. For example, Ando et al. (2013) showed that the critically endangered red-headed wood pigeon, Columba janthina nitens, consumed introduced plants more frequently than native species on the Ogasawara Islands, Japan. Similarly, introduced species were consumed frequently by the Ogasawara buzzard, Buteo buteo oyoshim, with 90% of its diet consisting of introduced animals (Kato & Suzuki, 2005). These small oceanic islands harbor high levels of endemism. Introduced species are typically associated with disproportionately negative effects on island biodiversity (Sax & Gaines, 2009), but are shown to provide nutritional resources to these endemic species. This may be more common than currently acknowledged, with introduced species representing a significant dietary element for native consumers.

Round Island, situated 22.5 km North-East of Mauritius (Figure 1), is a globally significant site of biodiversity conservation and now represents the last remnant of native lowland palm habitat (Figure 2) in the Mascarenes (Cheke & Hume, 2008). The palm habitat has been recovering since the eradication of goats, Capra aegagrus hircus, in 1979, and rabbits, Oryctolagus cuniculus, in 1986 (Cheke & Hume, 2008; Merton, 1987). At just 2.19 km², it is home to several reptile species extirpated from mainland Mauritius by introduced species and habitat destruction. Telfair’s skinks, Leiolopisma telfairii (Figure 3), are vulnerable omnivorous reptiles, typically growing to approximately 30 cm in total length, and are endemic to Mauritius. They became restricted to Round Island by the mid-1800s because of the introduction of non-native predators, such as rats (Cole, Goder, et al., 2018). The species has now been re-introduced to the island Nature Reserves, Ile aux Aigrettes (0.26 km², located 600 m from South-East Mauritius), and Gunner’s Quoin (0.7 km², 5 km to the North of Mauritius) (Cole, Goder, et al., 2018). Round Island has been a designated nature reserve since 1957 and has never suffered from introduced terrestrial vertebrate predators, which have caused the extirpation and extinction of multiple Mauritian species.
Habitat restoration efforts on Round Island since the 1980s have led to the recovery of its reptile populations, which includes seven species, four of which became restricted to the island by the mid-19th century (Cole, Mootooocurpen, et al., 2018; North et al., 1994).

Previous dietary analyses of Telfair’s skinks include morphological identification of food items and molecular analyses (Brown et al., 2014; Moorhouse-Gann et al., 2021; Pernetta et al., 2005; Zuël, 2009). Morphological examination of feces shows that Telfair’s skinks consume a variety of introduced and native species of fruit, seeds, arthropods, and vertebrates (Pernetta et al., 2005; Zuël, 2009). However, morphological methods of diet analysis can be unreliable and taxonomically imprecise, even when researchers are skilled. These methods also fail to adequately detect small or soft-bodied prey (Pompanon et al., 2012; Symondson, 2002). Previous HTS-based fecal analysis of Telfair’s skinks targeting plant (Moorhouse-Gann et al., 2021) and animal (Brown et al., 2014) food resources on Ile aux Aigrettes and Round Island confirmed skinks eat a diverse range of taxa.

Identifying the diet of omnivores is challenging, but a few studies have facilitated the most comprehensive complex dietary assessments to date using DNA metabarcoding (Bonin et al., 2020; De Barba et al., 2014; Robeson et al., 2018; Silva et al., 2019). Trophic generalists may be central to ecological networks and can elicit top-down effects across their entire breadth and depth. Deciphering the structure and dynamics of these interactions is therefore valuable, especially within a conservation context. Telfair’s skinks are large, locally abundant trophic generalists endemic to Mauritius (Cole, Goder, et al., 2018; Jones, 1993; Vinson & Vinson, 1969), and are therefore likely to exert strong top-down pressures on the ecological network of Round Island.

Here, we aimed to study the complete diet of Telfair’s skinks on Round Island by using broad-coverage plant and animal DNA metabarcoding primers. In doing so, we aimed to show: (a) the dietary diversity and key trophic interactions of Telfair’s skinks; (b) whether introduced species feature prominently in the diet; (c) whether diet composition changes between seasons; and (d) whether diet composition is different between male and female skinks, which may have implications for conservation management and reintroduction initiatives.

2 METHODS

2.1 Study site

Round Island (Figure 1) is a basaltic cone that reaches 280 m above sea level and retains the last remnant of a native lowland palm habitat within the Mascarenes (Cheke & Hume, 2008), which has
been recovering since introduced vertebrate herbivores were eradicated (Cheke & Hume, 2008; Merton, 1987). This habitat is primarily dominated by the blue latan palm, *Lattania loddigesii*. Introduced herbaceous plants, such as *Achyranthes aspera* and *Tridax procumbens*, form swathes of invaded habitat in large open clearings between thickets of native trees. Before the 1980s, much of the island suffered deforestation caused by the introduced herbivores, resulting in the loss of all but two hardwood tree species represented by a single individual bois buis, *Fernelia buxifolia*, tree and a few individuals of acacia indigène, *Gagnebina pterocarpa* (Strahm, 1993). Loss of habitat led to extensive soil erosion and created large expanses of barren rock slab over much of the island (Figure 2). Since 2002, there have been extensive efforts to restore the lost hardwood forests and to enhance the natural regeneration of the palm habitat (Jones, 2008). The invertebrate community is poorly studied, with few native species formally identified and described (Moldowan et al., 2016). Several introduced invertebrate species are now established over much of the island, such as the ants *Pheidole megacephala* and *Brachymyrmex cordemoyi*, and the webspinner *Oligotoma saundersii*. Endemic species of arthropod are common in their favored habitats, such as the Round Island stick insect, *Apterograeffea marshallae*, a herbivore of *L. loddigesii* (Moldowan et al., 2016), and the Serpent Island centipede, *Scolopendra abnormis*, a large invertebrate predator (Senapathi et al., 2010).

The vertebrate community consists of regionally important seabird colonies, a remnant endemic reptile assemblage and two introduced land bird species (Cheke & Hume, 2008; Cole, Mootoocurpen, et al., 2018). Seven endemic reptile species survive on Round Island because of the absence of introduced predators. Five of these are listed as Threatened on the IUCN Red List: Bojer’s skink, *Gongylomorphus bojerii*, Durrell’s Night gecko, *Nactus durrellorum*, keel-scaled boa, *Casarea dussumieri*, Round Island day gecko, *Phelsuma guentheri*, and Telfair’s skink, *Leiolopisma telfairii* (IUCN, 2020). Additionally, two tortoise species, Aldabra giant tortoise, *Aldabrachelys gigantea*, and radiated tortoise, *Astrochelys radiata*, have been introduced to Round Island as “ecological replacements” for extinct Mauritian tortoises, *Cylindraspis spp.* (Griffiths et al., 2010). As the largest and one of the most abundant of the island’s lizards, Telfair’s skinks constitute the largest component of animal biomass of any omnivore on the island (Cole, Mootoocurpen, et al., 2018) and likely have a significant role within the island’s food web dynamics.

Broad dry and wet seasons exist in Mauritius (Senapathi et al., 2010). The dry season typically begins in May and is characterized primarily by low rainfall, mean air temperature of ~20.5°C, and stronger winds, with the driest months being September and October. The wet season typically begins in December and is characterized by much more frequent rainfall, mean air temperature of ~24.5°C, and minimal winds, with the wettest months being January and February (Senapathi et al., 2010).

### 2.2 Skink sampling on Round Island

Fecal samples were collected in March, June, July, and December 2015 (Figure 1). Skinks were caught opportunistically by noose or hand after which defecation was induced using a gentle abdominal massage. The fecal samples were placed in polythene bags and dried over silica gel. Telfair’s skinks are present over the entire island, but, unfortunately, some areas of the island are too dangerous to capture these fast-moving reptiles. Skinks were released unharmed within ten minutes of capture at the locations where they were caught. Fecal samples were collected from 196 individual Telfair’s skinks (identified by their sex, size that were recorded and distinguishing markings and body deformations, which were photographed) on Round Island and previously underwent DNA metabarcoding to identify the floral component of skink diet (Moorhouse-Gann, 2018; Moorhouse-Gann et al., 2021). Due to funding constraints, we were only able to advance 82 samples to sequencing, which were randomly selected for the current study from both dry (40) and wet (42) seasons.

### 2.3 Primer selection

Animal primers were tested in silico with a broad range of vertebrate and invertebrate taxa using PrimerMiner (Elbrecht & Leese, 2017) and in vitro with DNA extracted from animals sampled on Round Island. BerenF-LuthienR (Cuff et al., 2020) provided the most comprehensive coverage, amplifying all Round Island invertebrate DNA extracts tested. UniPlant primer pair (Moorhouse-Gann et al., 2018) was used to amplify the ITS2 DNA barcode in plants and successfully amplify almost all plant species found on Round Island.

### 2.4 DNA extraction, PCR amplification, and sequencing

DNA extraction from Telfair’s skink fecal samples and preparation of plant DNA for 250-bp paired-end Illumina MiSeq high-throughput sequencing followed Moorhouse-Gann et al. (2021; Supplementary Information S2 and Table S1).

We used the following procedure to identify animal prey in the diet of Telfair’s skinks. Polymerase chain reactions (PCRs) used 25 μl reaction volumes containing 5 μl DNA template, 12.5 μl of multiplex PCR mix (Qiagen, Manchester, UK), 2.5 μl of both forward and reverse primers (0.2 μM each), and 2.5 μl of nuclelease-free water (Qiagen). PCR conditions are as follows: 95°C for 15 min, 35 cycles of 95°C for 30 s, 54°C for 90 s, and 72°C for 10 min, as instructed by the manufacturer (Qiagen). Each sample incorporated a unique combination of molecular identification (MID) tags (Binladen et al., 2007) that allowed for each skink to be identified after pooling and sequencing as per Brown et al. (2014). These 10-bp fragments were added to both the forward and reverse primers for each sample, and thus, dietary taxon sequences could
be assigned to individuals. PCR products were then run through a 2% agarose gel stained with SYBR®Safe (Thermo Fisher Scientific, Paisley, UK). Twelve negatives were included in each PCR run, 10 PCR negatives, and two extraction negatives. Additionally, two positive controls consisting of a standardized DNA concentration (4 ng/μl) of known invertebrate species likely absent from the study site (Supplementary Information S1) were used to control for tag-jumping between samples in the filtering steps detailed below. PCR products were run in a Qiagen QIAXcel Advanced System (Qiagen) to measure relative DNA concentrations and later measured individually using a Qubit Fluorometer (Thermo Fisher Scientific) for more accurate determination of DNA concentrations. Each sample was then pooled based on the relative DNA concentrations of the amplicon of interest as measured by the QIAXcel Advanced System. Negative controls were pooled based on the average volume pooled for the skim samples. The pooling process involved adding a volume from each sample as a proportion of the sample with the highest concentration of DNA, to ensure approximate equimolarity of DNA from each sample. Each pool was cleaned using SPRINTselect beads (Beckman Coulter, Brea, USA), with a left-side size selection using a 1:1 ratio. After final elution, the pool was run on a Qubit Fluorometer, to measure DNA concentration (=49.6 ng/μl), as well as an Agilent 2200 TapeStation with D1000 ScreenTape (Agilent Technologies, Waldbronn) to check for significant levels of primer dimer, which were not found. This pool of MID-tagged samples was then used for library preparation using the NEXTflex™ Rapid DNA-Seq Kit following the manufacturer’s instructions (Bioo Scientific Corp, Austin, TX, USA), which is suitable for pools with DNA concentrations of 1 ng–1 μg. A final DNA concentration was measured for the prepared library using a Qubit Fluorometer (=11.7 ng/μl) and was then sequenced on an Illumina MiSeq desktop sequencer (Illumina, San Diego, CA, USA) with a Nano cartridge using 2 × 250 bp paired reads (expected reads ≤1,000,000).

2.5 | Bioinformatics

The Illumina Nano cartridge run generated 750,645 reads. High-throughput sequencing data for the animal component of Telfair’s skink diet followed the bioinformatic process of Drake et al. (2021): FastP (Chen et al., 2018) was used to check the quality of reads, discard poor quality reads (<Q30, <125 bp long or too many unqualified bases, denoted by “N”), trim reads to a minimum length of 300 bp and merge read pairs from Miseq files (R1 and R2). Read pairs were assigned to samples and demultiplexed using Mothur v1.39.5 (Schloss et al., 2009), after which MID-tag and primer ends were removed. Unoise3 (Edgar, 2010) was used to remove replicates, denoise the sequences, and group identical sequences into zero-radius operational taxonomic units (ZOTUs), which are clustered without % identity to avoid multiple species being nested within an OTU. Processed sequences were given taxonomic information from GenBank using BLASTn v2.7.1 (Camacho et al., 2009) with a 93% identity threshold. This threshold was chosen to capture the wide variety of invertebrates on Round Island to genus- or family-level, most of which have not been barcoded or formally described. When more than one taxon was assigned to a sequence, we manually checked the feasibility for the presence of each taxon on Round Island by searching published articles, unpublished reports, and personal observations of species accounts. If these manual checks were inconclusive, we assigned the sequence to a higher taxonomic level (genus, family, order, etc.). MEGAN Community Edition v6.18.9 (Huson et al., 2016) was used to analyze the BLAST output and assign taxonomic identities to each ZOTU. Using the lowest e-value (a value estimating the number of hits “expected” by chance when searching a database of a given size—in this instance anything <0.00001), the top hit was assigned to each sequence. Where top hits were taxonomic levels higher than species, these were manually checked and assigned to a feasible taxon or deleted from the analysis if erroneous. ZOTUs that were assigned to the same taxon were aggregated.

Data were cleaned for statistical analysis following the methods set out by Drake et al. (2021): The combined removal of the maximum read count in blanks and negative controls, and reads not meeting a predefined per sample threshold, removes both erroneous reads (laboratory contaminants and sequencing errors) that are likely to occur in low abundances mitigate tag-jumping and bleeding of over-represented taxa into other samples, while utilizing a per sample threshold and those arising through tag-jumping and bleeding of over-represented taxa into other samples removes erroneous reads (laboratory contaminants and sequencing errors) that are likely to occur in low abundances. The maximum read count of known contaminants and other obviously erroneous ZOTUs across the dataset was calculated as a percentage of their respective total sample read count, and any read counts less than this were removed. For this, a threshold of 0.3% was applied, removing low-frequency laboratory contaminants and sequencing errors. Following this, the highest read count within a blank or negative per ZOTU was calculated and any ZOTU reads below this value were removed. In addition, we established an extra per-ZOTU filtering step, which removed remaining erroneous taxa. The per-ZOTU threshold was set to 0.74%. After these filters were applied, read counts were converted to presence-absence data for each sample. Nine samples were removed due to the absence of any dietary detections, leaving 73 samples to be taken forward for statistical analyses. Bioinformatic analysis for plant sequencing data followed Moorhouse-Gann et al. (2021) (Supplementary Information S2).

After animal ZOTUs were given taxonomic information, status of each taxon relative to Round Island was determined for each by manually searching for relevant data in published articles, unpublished reports, and personal species accounts, and then classified as “cryptogenic,” “endemic,” “introduced,” or “native.” Cryptogenic species were defined as species that had no clear status, either because of poor taxonomic resolution, or because they may be known natives of the Indian Ocean islands, but their history on Round Island is unknown. Plant status was taken from Moorhouse-Gann (2018; Moorhouse-Gann et al., 2021).
2.6 | Statistical analyses

Statistical analyses were conducted in R Statistical Software v4.1.0 (R Core Team, 2021) after data were converted to presence/absence within each sample. Basic characteristics of the diet were quantified by measuring frequency of occurrence. We aimed to reveal whether there were significant differences in the mean frequency of occurrence of dietary taxa from different taxonomic kingdoms (animals, plants) or status relative to Round Island (cryptic, endemic, introduced, native), hereafter "status." Data were not normally distributed (Shapiro–Wilk test for normality: $W = 0.64, p = <.001$), and we therefore used two nonparametric Kruskal–Wallis tests, one each for kingdom and status, to determine whether there were significant differences in average consumption between categories of each variable.

We also wanted to quantify dietary diversity and show whether our samples could be used to sufficiently represent the broad dietary patterns of Telfair's skinks. Sample size and effort-based standardization poorly represent the true diversity of communities because they fail to account for the species-abundance distribution of the community being sampled (Cao et al., 2007; Roswell et al., 2021). We therefore used coverage-based rarefaction and extrapolation rather than asymptotic species-accumulation curves (Chao & Jost, 2012; Roswell et al., 2021) and robustly estimated species diversity using Hill diversity (Hill, 1973; Roswell et al., 2021). We define Hill diversity by the equation,

$$D = \left( \sum_{i=1}^{S} p_i \left( r_i \right)^i \right)^{1/i}$$  \hspace{1cm} (1)

where $D$ is diversity, $S$ is number of species, $p_i$ is the proportion of all individuals that belong to species $i$, $r_i$ is the rarity of species $i$, defined as $1/p_i$, and $i$ is the exponent determining the rarity scale on which the mean is taken (Bullen, 2003; Hill, 1973; Roswell et al., 2021). Hill diversity is the generalized mean species rarity, and the exponent $i$ determines the sensitivity of the equation to rare species. $i$ of 1 uses the arithmetic mean rarity, or species richness (Hill-richness), and is very sensitive to the rarest species; $i$ of 0 uses the geometric mean rarity, or the exponential of Shannon's entropy (Hill-Shannon), and responds to both high and low rarity species; and $i$ of -1 uses the harmonic mean rarity, or the inverse of Simpson's index (Hill-Simpson), and is most sensitive to the relative abundance of common species (Roswell et al., 2021). Coverage is a measure of how completely a community has been sampled and is an estimated proportion of the sampled individuals in the community that belong to species already detected (Chao & Jost, 2012). For example, a coverage of 0.88 denotes that 15% of the individuals in the community being sampled belong to species that have not been found. We computed these metrics in R package "iNEXT" (Hsieh et al., 2016).

Variation in diet composition was visualized with nonmetric multidimensional scaling (NMDS) in the "vegan" package (Oksanen et al., 2019) using the "metaMDS" function on a matrix of Jaccard distances, where we extracted three dimensions. Data were plotted using package "ggplot2" (Valero-Mora, 2010). To illuminate whether sex, season, or their interaction affects Telfair's skink diet, R package "mvabund" was used (Wang et al., 2012). Multivariate generalized linear models (MGLMs) were run using the "manyglm" function with a Monte Carlo resampling method and "binomial" error family. The "step" function facilitated model selection where we selected the lowest AIC value denoting which model was most supported given the data.

3 | RESULTS

There were 389 dietary presence counts belonging to 77 dietary taxa found across the 73 Telfair's skinks samples. Of these, 37 of 38 plant taxa were resolved to species due to extensive barcoding of the Round Island flora. The invertebrates of Round Island have not been described as extensively, and of the 39 dietary taxa detected, 20 were resolved to species, nine to genus, nine to family, and one to order. The invasive ant P. megacephala and a cryptogenic braconid wasp, Heterosphilus sp., were the most frequently detected taxa, present in almost 40% of all Telfair's skink samples (Table 1; Table S2).

Our Kruskal–Wallis tests showed that mean number of detections per dietary taxon was not significantly affected by status relative to Round Island ($\chi^2(3) = 1.51, p = .68$), but taxonomic kingdom did show a significant effect ($\chi^2(1) = 6.33, p = .012$), where plants were consumed more frequently on average per taxon than animals (mean consumption per dietary taxon (± SE): animals = 4.26 (± 1.2), plants = 5.87 (± 1.03); Figure 4, Table S3). Introduced taxa were cumulatively consumed more frequently than all other status categories, accounting for 49.4% of all detections, while cryptogenic and native taxa accounted for 20.6% each, and endemic taxa 9.5% (Figure 5).

We computed dietary diversity (Figure 6) and found that Hill-richness ($i = 1$) provided the highest diversity estimate in contrast to both Hill-Shannon ($i = 0$) and Hill-Simpson ($i = -1$) (Figure 6, left plot). Together, these diversity estimates suggest Telfair's skinks consume many rarely eaten individual species instead of evenly consuming dietary taxa or just a few commonly eaten species. We estimated that our sampling provided 95.7% (±95% CI: 2.6%) coverage of the dietary community (Figure 6, center and right plots), meaning that 4.3% of individuals in the theoretical diet belonged to species we did not detect.

MGLMs showed that diet composition differed significantly between season (Wald = 259.88, $p = <.001$), sex (Wald = 226.22, $p = <.001$), and their interaction (Wald = 30.54, $p = .027$). Diet composition between seasons was visualized using NMDS (Figure 7) (stress = 0.161). Three species showed at least one significant GLM result: Abutilon indicum (season*sex: Wald = 9.035, $p = .031$), A. aspera (season: Wald = 31.097, $p = <.001$), and L. loddigesii (sex: Wald = 25.161, $p = .002$; season: Wald = 31.213, $p = <.001$).
**TABLE 1** Taxonomic information, frequency of occurrence $F_o$ (%), and status relative to Round Island (cryptogenic, endemic, introduced, native) for all dietary taxa occurring in two or more Telfair's skink fecal samples

| Kingdom     | Phylum     | Class       | Order     | Family       | Dietary taxon                  | $F_o$ (%) | Status             |
|-------------|------------|-------------|-----------|--------------|--------------------------------|-----------|--------------------|
| Animalia    | Arthropoda | Arachnida   | Araneae   | Thomisidae   | Ozyptila claveata              | 5.48      | Introduced         |
| Crustacea   | Isopoda    | Porcellionida | Blaberida | Blaberidae sp. | Porcellionidae sp.            | 34.25     | Cryptogenic       |
| Insecta     | Blattodea  | Coccinellida | Drosophilida | Drosophila melanogaster | | 2.74 | Cryptogenic |
| Coleoptera  | Diptera    | Tachinidae  | Chetogena | Chetogena sp. | | 2.74 | Cryptogenic |
| Embioptera  | Oligotomida | Insecta    | Oligotoma | Oligotoma saundersii | | 2.74 | Introduced |
| Hemiptera   | Aleurodida | Hymenoptera | Apidae    | Inquiline | | 2.74 | Native |
|             |            | Rhyparochromida | Rhyparochromida | Rhyparochromidae | | 8.22 | Cryptogenic |
|             |            | Hymenoptera | Braconidae | Heterospilus sp. | | 39.73 | Cryptogenic |
|             |            | Formicidae  | Brachymyrmex cordemoyi | Formicidae sp. | | 19.18 | Introduced |
|             |            |              | Monomorium | Monomorium florica | | 2.74 | Cryptogenic |
|             |            | Hymenoptera | Heterospilus | Heterospilus sp. | | 39.73 | Cryptogenic |
|             |            |              | Boerhavia | Boerhavia sp. | | 19.18 | Native |
|             |            |              | Gymnosporia | Gymnosporia pyria | | 2.74 | Endemic |
|             |            |              | Fabaceae  | Desmodium incanum | | 6.85 | Introduced |
|             |            |              | Fabaceae  | Gognobia ptercarpa | | 10.96 | Native |
|             |            |              | Apocynaceae | Apocynaceae | | 10.96 | Native |
|             |            |              | Lamiales  | Premna serratifolia | | 5.48 | Native |
|             |            |              | Euphorbiaceae | Euphorbia thymifolia | | 4.11 | Cryptogenic |
|             |            |              | Passifloraceae | Passiflora suberosa | | 15.07 | Introduced |
|             |            |              | Phylanthaceae | Margaritaria anomal | | 4.11 | Endemic |
|             |            |              | Malvaceae | Abutilon indicum | | 35.62 | Introduced |
|             |            |              | Malvaceae | Hibiscus tiliaeus | | 4.11 | Native |
|             |            |              | Myrtaceae | Eugenia lucida | | 4.11 | Endemic |
|             |            |              | Solanaceae | Ipomoea pes-caprae | | 21.92 | Native |
|             |            |              | Solanaceae | Solanum lycopersicum | | 4.11 | Introduced |
|             |            |              | Monocots | Solanum nigrum | | 17.81 | Introduced |
|             |            |              | Poales | Latania loddigesii | | 32.88 | Endemic |
|             |            |              | Poaceae | Solanum echinatus | | 13.70 | Introduced |
|             |            |              | Poaceae | Chloris barbata | | 2.74 | Introduced |
|             |            |              | Poaceae | Digitaria horizontalis | | 4.11 | Native |
|             |            |              | Poaceae | Digitaria horizontalis | | 12.33 | Introduced |

Note: Cryptogenic taxa had no clear status, either because of poor taxonomic resolution, or because they may be known natives of the Indian Ocean islands, but their history on Round Island is unknown.
4 | DISCUSSION

4.1 | Key trophic interactions and dietary diversity

These findings corroborate previous analyses of diet, suggesting Telfair’s skinks are generalist omnivores that consume a wide range of animal and plant taxa (Brown et al., 2014; Cole, Goder, et al., 2018; Moorhouse-Gann et al., 2021; Pernetta et al., 2005). Moreover, our diversity estimates suggest Telfair’s skinks consume many species infrequently instead of consuming taxa evenly. This study achieved a greater taxonomic resolution compared to previous molecular analyses of Telfair’s skink diet (Brown et al., 2014; Pernetta et al., 2005), resolving almost all plant taxa and nearly half of the invertebrate taxa to species-level. In contrast, previous analyses could not resolve dietary invertebrate taxa to species level at all.

Dietary taxa consumed once or twice form a large component of Telfair’s skink diet, suggesting that they may opportunistically consume many rare species, but rely on a few other species for more consistent nutrition, which may also be seasonal. For example, *L. loddigesii* was the most frequently consumed native dietary taxon, being present in 32.9% of samples, and all detections occurred in the dry season. These trees form the dominant native habitat type on Round Island and produce fruits, pollen, and nectar that skinks are known to readily consume (Cole, Goder, et al., 2018; Cole, Mootooocurpen, et al., 2018). Since all detections of *L. loddigesii* occurred in the dry season, when fruits are produced, this suggests they are a seasonal nutritional resource for Telfair’s skinks. Moreover, female skinks accounted for 60% of all *L. loddigesii* detections and fruits may therefore be disproportionately important or attractive to female skinks. Telfair’s skinks typically mate throughout the dry season (Cole, Goder, et al., 2018) and...
L. loddigesii consumption may provide essential nutrition or minerals for growth and/or egg production in females. In contrast, 34 dietary taxa were detected only once, 12 taxa were detected twice, and six taxa were detected three times, cumulatively representing almost a fifth of total detections. Our diversity estimates suggest Telfair’s skinks consume many dietary taxa at relatively low frequencies. This may represent the foraging behavior of Telfair’s skinks whereby they feed on a few species regularly but supplement their diet by opportunistically consuming a much greater diversity of other animals and plants, albeit at a lower frequency for each taxon.

There are an estimated 46,000 Telfair’s skinks on Round Island (Cole, Mootoocurpen, et al., 2018), with an estimated 210 skinks per ha island-wide. This represents a major component of total animal biomass. Given the abundance and size of Telfair’s skinks, these results show they are likely to represent a major top-down pressure on the ecological network through their dietary generalism.

4.2 | Prevalence of introduced taxa

Overall, introduced taxa formed the primary component of Telfair’s skink diet as measured by frequency of occurrence. The majority of dietary detections and richness were of introduced taxa, accounting for almost half in both cases. Therefore, this study illuminates that introduced taxa have become a large part of the diet of a globally threatened endemic species. However, for some taxa it is unclear whether skinks rely on them for nutrition, and this is a broader issue in dietary metabarcoding studies because sequencing data cannot convey nutritional information (Alberdi et al., 2019; Lamb et al., 2019). For example, the introduced ant P. megacephala is present in 39.7% of samples, but may be a distasteful meal for Telfair’s skinks. On Round Island, P. megacephala is hyperabundant and found in every habitat type in this study. Predation may not provide a cost-effective nutritional reward to an unspecialized ant-eating vertebrate given that the ant is very small compared to Telfair’s skinks.
and ants typically possess unpleasant and/or harmful compounds (Schmidt, 2009). If these ants truly are deleterious to skinks, their high frequency of occurrence in the diet could be explained by accidental consumption. Accidental consumption may occur when skinks consume food items that have been colonized by ants, which typically occurs rapidly on Round Island. Another explanation is through secondary predation, which entails detection of food items in the digestive system of primary skink prey. Both accidental consumption and secondary predation may complicate interpretation of dietary analyses using HTS (Robeson et al., 2018; Silva et al., 2019; Tercel et al., 2021). Nevertheless, even accidental ingestion of some species could provide nutritional benefits to skinks.

With roughly half of all dietary detections originating from introduced species, non-native taxa appear to be a dominant part of Telfair’s skink diet. It may be that the original components of the diet have been lost after Round Island suffered severe habitat destruction and have been subsequently replaced by non-native species. Equally, the availability or nutritional value of non-native species may be relatively higher than existing native food.

Cryptogenic invertebrates represented almost 20% of all dietary detections and therefore likely represent an important component of Telfair’s skink diet. In reality, cryptogenic species are either introduced or native, but this information is lost without adequate taxonomic information. Unfortunately, many of the invertebrates on Round Island remain undescribed and are absent from barcode reference libraries, which presents a problem when assigning an origin to ZOTUs that do not resolve to species-level. Furthermore, many undescribed invertebrate species categorized as cryptogenic may be endemic and globally threatened. The use of a 93% identity threshold in this study permits assignment of sequences to a higher taxonomic level for species absent from barcode libraries, that is, to genus or family, but this does not solve how to assign a dietary taxon a status. Our study deliberately took a conservative approach to assigning a status category to taxa, but it may be more likely for cryptogenic species to be native than introduced. This is because many introduced species are globally common and have been barcoded, whereas endemic species have not. Work to formally describe, identify, and barcode Round Island invertebrate species is therefore essential to disentangling this problem and to more fully describing the ecology of Round Island.

4.3 | Seasonal and sex differences

The presence of plant species in the diet of the skinks solely or primarily in one season, such as L. lodgesi (all 24 detections in the dry season) and A. aspera (all 14 detections in the wet season), partly explains the strong difference in diet composition between seasons. Broad seasonal differences in diet are further explained by animal taxa being a much greater component of skink diet in the dry season, where 61.4% of animal prey detections occurred. Despite this, most animal taxa were consumed across both seasons, while most plant taxa were consumed primarily in only one season. This confirms that Telfair’s skinks rely on different dietary taxa at different times of the year at a broad scale, with only modest overlap in composition (Figure 7). Seasonal differences in diet composition very likely arise because of changes in the availability of food sources between the markedly different seasons in Mauritius (Senapathi et al., 2010).

Two dietary taxa were consumed at different rates between male and female skinks: L. lodgesi, as discussed above, and A. indica, which was the most frequently consumed plant, present in 35.6% of samples. Abutilon indica is native to tropical and subtropical Asia but has been widely transported across the global tropics and is locally abundant over much of Round Island. It produces flowers and seeds that may be attractive to skinks year-round. Male skinks consumed A. indica more frequently in the wet season (71.4% of detections) than the dry season (28.6%), but the inverse was true of female skinks (100% of detections in the dry season). A possible explanation is that plant tissues of A. indica (e.g., flower, nectar, seeds) are consumed differently between sexes. Because the availability of these varies throughout the year, it may mean that males and females consume A. indica differently between seasons. Exactly how the tissue types of A. indica may differentially benefit male and female skinks requires further study.

Understanding the nutritional requirements between sexes could be an important factor governing the success of skink translocations. Although we broadly see that male and female skinks consume the same species, we show that female skinks might rely more on certain species during the breeding season, which is a pivotal period in any reintroduction program.

4.4 | Limitations

The general limitations of dietary metabarcoding have been reviewed extensively by other authors (Alberdi et al., 2019; Lamb et al., 2019; Nielsen et al., 2018; Taberlet et al., 2018), but we also identified some study-specific limitations. This study converts sequence data to presence/absence and subsequently frequency of occurrence. We believe this is the most robust interpretation of sequencing data, because sequencing output only very weakly correlates with biomass in a sample (Deagle et al., 2019; Lamb et al., 2019). Nevertheless, frequency of occurrence therefore also omits how much biomass is consumed in each sample and, thus, a dietary taxon may appear frequently between samples but not contribute proportionately to the nutrition of the consumer.

As discussed above, the very high prevalence of introduced ants in Telfair’s skink diet is difficult to explain ecologically with any certainty. These have not been observed to be directly eaten by the skinks, but are ubiquitous over Round Island, and colonize food resources rapidly. Moreover, a very frequently found tiny (<2 mm) cryptogenic braconid wasp, Heterospilus sp., seems unlikely to be actively preyyed upon by adult Telfair’s skinks. Accidental consumption or secondary predation might explain these detections, as has been seen in other dietary metabarcoding studies (Silva et al., 2019), and have been identified as a potential source of error that
may disproportionately complicate the interpretation of dietary analyses of omnivores (Tercel et al., 2021). With an aim to tease apart some of these issues, we conducted a co-occurrence analysis (Supplementary Information 3; Figure S1) but found no clear ecological patterns that explain these detections. Indeed, co-occurrence analyses may be used as an exploratory element in ecological studies but cannot provide strong evidence to support ecological hypotheses in this context (Blanchet et al., 2020), and may not facilitate interpretation (Tercel et al., 2021).

Omnivores can exert top-down effects across the breadth and depth of ecological networks, and studying their diet is therefore valuable to the field of ecology. However, omnivorous diets require extra caution when inferring ecological conclusions from sequencing results given that some detections may not be ecologically meaningful. In this study, it may be that Telfair’s skinks are directly consuming both *P. megacephala* and *Heterosplius* sp., but it remains unclear whether this is true from conflicting behavioral observations of Telfair’s skinks and our inconclusive co-occurrence analysis that does not provide alternative ecological explanations.

### 4.5 Concluding remarks

Our study represents one of only a few complete dietary analyses of omnivores using DNA metabarcoding (but see De Barba et al., 2014; Ducotterd et al., 2021; Robeson et al., 2018; Silva et al., 2019) and the first study examining the omnivorous diet of a threatened endemic reptile. We found that Telfair’s skinks consume a few species regularly and many species rarely. We also found that Telfair’s skinks rely on *L. loddigesii* fruits during the dry season on Round Island, coinciding with when breeding takes place. Though restored habitat on Round Island does not cover the whole island, extensive habitat regeneration efforts since 2002 have led to the continued recovery of the forests on Round Island, and this bodes well for the future of the skinks. Nevertheless, almost half of all dietary detections were of introduced species and it is increasingly clear that the ecological impacts of introduced species on Round Island are multifaceted, with some species acting as *de facto* ecological replacements. Telfair’s skinks have probably incorporated introduced species into their diet as a replacement for the many species that were presumably lost by severe habitat destruction on Round Island. Further studies showing the mechanisms by which Telfair’s skinks consume certain species would help to provide more confident ecological explanations for some of these interactions.

Although many introduced species damage native ecosystems, some species may benefit regenerating habitats and their inhabitants, particularly in the context of island restoration. This study shows that many introduced species of animal and plant contribute positively to providing nutritional subsidies to a globally threatened endemic omnivore. Positive effects of introduced species must therefore be weighed up against potential negative consequences of colonization for the ecosystem. This is pertinent for conservation managers to consider when restoring native habitats and controlling introduced species, especially when threatened animal species may be consuming introduced taxa in the absence of lost native food resources.

### ACKNOWLEDGMENTS

MPTGT was supported by the British Herpetological Society (BHS Student Grant Scheme, ref: 517513), Durrell Wildlife Conservation Trust (MR/SS02455/1), and the Natural Environment Research Council (NE/L002434/1). Additionally, the authors would like to thank the Mauritius Wildlife Foundation for in-kind support and the National Parks and Conservation Service of Mauritius for permission to conduct research on Round Island. MPTGT extends further thanks to the Round Island Wardens, Johannes Chambron and Bethan Govier, for their expert advice on the biology of Telfair’s skinks and wider Round Island ecosystem. RJMG was funded by a NERC iCASE studentship (NE/K00719X/1), co-funded by CASE partners the Durrell Wildlife Conservation Trust with support from the Mauritian Wildlife Foundation. Thank you to the interns and volunteers who assisted with sampling. Illumina library preparation, sequencing, and data analyses were supported by the NERC Biomolecular Analysis Facility at the University of Sheffield, and funded by NERC, UK (NBAF983). We thank Gavin Horsburgh for assistance with library preparation, which was sequenced on an Illumina MiSeq by the Sheffield Diagnostic Genetic Service, part of the Sheffield Children’s NHS Foundation Trust. The authors would like to thank two anonymous reviewers for their excellent comments that helped improve the manuscript.

### CONFLICT OF INTEREST

The authors declare there are no conflicts of interest.

### AUTHOR CONTRIBUTION

Maximillian P. T. G. Tercel: Data curation (equal); Formal analysis (lead); Investigation (lead); Methodology (equal); Project administration (lead); Resources (equal); Software (equal); Validation (equal); Visualization (equal); Writing – original draft (lead); Writing – review & editing (lead). Rosemary J. Moorhouse-Gann: Data curation (equal); Funding acquisition (lead); Investigation (equal); Methodology (equal); Project administration (supporting); Writing – review & editing (supporting). Jordan P. Cuff: Formal analysis (equal); Investigation (equal); Methodology (equal); Writing – review & editing (supporting). Lorna E. Drake: Formal analysis (equal); Investigation (equal); Methodology (equal); Writing – review & editing (supporting). Nik C. Cole: Conceptualization (equal); Funding acquisition (equal); Investigation (equal); Methodology (equal); Project administration (equal); Supervision (equal); Writing – review & editing (equal). Martine Goder: Investigation (equal); Methodology (equal); Project administration (equal). Rouben Mootoo Curpen: Investigation (equal); Methodology (equal); Project administration (supporting). William O. C. Symondson: Conceptualization (equal); Funding acquisition (lead); Methodology (equal); Project administration (lead); Supervision (lead); Writing – review & editing (equal).
DATA AVAILABILITY STATEMENT
Animal sequencing data are available from Zenodo: https://zenodo.org/record/5476197#.YbHafNDP02y. Plant sequencing data are deposited in the SCBI Sequence Read Archive.

ORCID
Maximillian P. T. G. Tercel ORCID: https://orcid.org/0000-0003-0820-3278
Jordan P. Cuff ORCID: https://orcid.org/0000-0002-0198-4940
Lorna E. Drake ORCID: https://orcid.org/0000-0003-0860-555X

REFERENCES
Alberdi, A., Aizpurua, O., Bohmann, K., Gopalakrishnan, S., Lynggaard, C., Nielsen, M., & Gilbert, M. T. P. (2019). Promises and pitfalls of using high-throughput sequencing for diet analysis. Molecular Ecology Resources, 19(2), 327–348. https://doi.org/10.1111/1755-0998.12960
Ando, H., Setsuko, S., Horikoshi, K., Suzuki, H., Umehara, S., Inoue-Murayama, M., & Isagi, Y. (2013). Diet analysis by next-generation sequencing indicates the frequent consumption of introduced plants by the critically endangered red-headed wood pigeon (Columba janthina nitens) in oceanic island habitats. Ecology and Evolution, 3(12), 4057–4069. https://doi.org/10.1002/ece3.773
Arbetman, M. P., Meeus, I., Morales, C. L., Alzen, M. A., & Smagghe, G. (2013). Alien parasite hitchhikes to Patagonia on invasive bumblebee. Biological Invasions, 15(3), 489–494. https://doi.org/10.1007/s10530-012-0311-0
Baldock, K. C. R., Goddard, M. A., Hicks, D. M., Kunin, W. E., Mitschunas, N., Osgathorpe, L. M., Potts, S. G., Robertson, K. M., Scott, A. V., Stone, G. N., Vaughan, I. P., & Memmott, J. (2015). Where is the UK’s pollinator biodiversity? The importance of urban areas for flower-visiting insects. Proceedings of the Royal Society B: Biological Sciences, 282(1803), 20142849.
Binladen, J., Gilbert, M. T. P., Bollback, J. P., Panitz, F., Bendixen, C., Nielsen, R., & Willerslev, E. (2007). The use of coded PCR primers enables high-throughput sequencing of multiple homolog amplification products by 454 parallel sequencing. PLoS One, 2(2), 1–9. https://doi.org/10.1371/journal.pone.0000197
Blanchet, F. G., Cazelles, K., & Gravel, D. (2020). Co-occurrence is not evidence of ecological interactions. Ecology Letters, 23(7), 1050–1063. https://doi.org/10.1111/ele.13525
Bonin, M., Dussault, C., Taillon, J., Lecomte, N., & Côté, S. D. (2020). Combining stable isotopes, morphological, and molecular analyses to reconstruct the diet of free-ranging consumers. Ecology and Evolution, 10(13), 6664–6676. https://doi.org/10.1002/2041-210X.12613
Brown, D. S., Burger, R., Cole, N., Vencatasamy, D., Clare, E. L., Montazam, A., & Symondson, W. O. C. (2014). Dietary competition between the alien Asian Musk Shrew (Suncus murinus) and a re-introduced population of Telfair’s Skink (Leiolepis telfairii). Molecular Ecology, 23(15), 3695–3705. https://doi.org/10.1111/mec.12445
Bullen, P. S. (2003). Handbook of means and their inequalities. Kluwer Academic Publishers.
Camacho, C., Coulouris, G., Avagyan, V., Ma, N., Papadopoulos, J., Bealer, K., & Madden, T. L. (2009). BLAST+: Architecture and applications. BMC Bioinformatics, 10, 1–9. https://doi.org/10.1186/1471-2105-10-421
Cao, Y., Hawkins, C. P., Larsen, D. P., & Van Sickle, J. (2007). Effects of sample standardization on mean species detectabilities and estimates of relative differences in species richness among assemblages. The American Naturalist, 170(3), 381–395. https://doi.org/10.1086/520117
Chao, A., & Jost, L. (2012). Coverage-based rarefaction and extrapolation: Standardizing samples by completeness rather than size. Ecology, 93(12), 2533–2547. https://doi.org/10.1890/11-1952.1
Cheke, A., & Hume, J. P. (2008). Lost land of the dodo: An ecological history of mauritius, réunion & rodrigues, 1st ed. Bloomsbury.
Chen, S., Zhou, Y., Chen, Y., & Gu, J. (2018). Fastsp: An ultra-fast all-in-one FASTQ preprocessor. Bioinformatics, 34(17), 1884–1890. https://doi.org/10.1093/bioinformatics/bty560
Clavero, M., & Garcia-Berthou, E. (2005). Invasive species are a leading cause of animal extinctions. Trends in Ecology and Evolution 20(3):110. TRENDS in Ecology and Evolution 19(1):17071
Cole, N., Goder, M., Premanand, R., Bachraz, V., & Mootooocurpen, R. (2018). Leiolepis telfairii. The IUCN Red List of Threatened Species, 2018, 8235.
Cole, N. C., & Harris, S. (2011). Environmentally-induced shifts in behavoiur intensify indirect competition by an invasive gecko in Mauritius. Biological Invasions, 13(9), 2063–2075. https://doi.org/10.1007/s10530-011-0225-8
Cole, N., Mootooocurpen, R., & Nundlial, V. (2018). Relative density estimates of Round Island’s reptiles. Journal of the Royal Society of Arts and Sciences of Mauritius 1(Vinson 1949):1–113.
Cuff, J. P., Drake, L. E., Tercel, M. P. T. G., Stockdale, J. E., Orozco-TorWengel, P., Bell, J. R., Vaughan, I. P., Müller, C. T., & Symondson, W. O. C. (2021). Money spider dietary choice in pre- and post-harvest cereal crops using metabarcoding. Ecological Entomology, 46(2), 249–261. https://doi.org/10.1111/een.12957
De Barba, M., Miquel, C., Boyer, F., Mercier, C., Rioux, D., & Taberlet, P. (2014). DNA metabarcoding multiplexing and validation of data accuracy for diet assessment: Application to omnivorous diet. Molecular Ecology Resources, 14(2), 306–323. https://doi.org/10.1111/1755-0998.12188
Deagle, B. E., Thomas, A. C., McInnes, J. C., Clarke, L. J., Vesterinen, E. J., Clarke, E. L., Kartzeniel, T. R., & Eveson, J. P. (2019). Counting with DNA in metabarcoding studies: How should we convert sequence reads to dietary data? Molecular Ecology, 28(2), 391–406. https://doi.org/10.1111/mec.14734
Drake, L. E., Cuff, J. P., Young, R. E., Marchbank, A., Chadwick, E. A., & Symondson, W. O. C. (2021). An assessment of minimum sequence copy thresholds for identifying and reducing the prevalence of artefacts in dietary metabarcoding data. Methods in Ecology and Evolution, 1–17. https://doi.org/10.1111/2041-210X.13780
Ducotterd, C., Crowadore, J., Lefort, F., Rubin, J., & Ursenbacher, S. (2021). A powerful long metabarcoding method for the determination of complex diets from faecal analysis of the European pond turtle (Emys orbicularis, L. 1758). Molecular Ecology Resources, 21(2):433–447. https://doi.org/10.1111/1755-0998.13277
Edgar, R. C. (2010). Search and clustering orders of magnitude faster than BLAST. Bioinformatics, 26(19), 2460–2461. https://doi.org/10.1093/bioinformatics/btq461
Elbrecht, V., & Leese, F. (2017). Development and validation of DNA metabarcoding COI primers for aquatic invertebrates using the R package ‘PrimerMiner’. Methods in Ecology and Evolution, 8(5), 622–626. https://doi.org/10.1111/peerj.preprints.2044v1
Griffiths, C. J., Jones, C. G., Hansen, D. M., Puttoo, M., Tatayah, R. V., Müller, C. B., & Harris, S. (2010). The use of extant non-indigenous tortoises as a restoration tool to replace extinct ecosystem engineers. Restoration Ecology, 18(1), 1–7. https://doi.org/10.1111/j.1526-100X.2009.00612.x
Hill, M. O. (1973). Diversity and evenness: A unifying notation and its consequences. Ecology, 54(2), 427–432. https://doi.org/10.2307/1934352
Hsieh, T. C., Ma, K. H., & Chao, A. (2016). iNEXT: an R package for rarefaction and extrapolation of species diversity (v ii numbers) McInerny, G. (ed.). Methods in Ecology and Evolution, 7(12):1451–1456. https://doi.org/10.1111/2041-210X.12613
Taberlet, P., Coissac, E., Hajibabaei, M., & Rieseberg, L. H. (2012). Environmental DNA. *Molecular Ecology*. https://doi.org/10.1111/j.1365-294X.2012.05542.x

Tercel, M. P. T. G., Symondson, W. O. C., & Cuff, J. P. (2021). The problem of omnivory: A synthesis on omnivory and DNA metabarcoding. *Molecular Ecology*, 30(10), 2199–2206. https://doi.org/10.1111/mec.15903

Valero-Mora, P. M. (2010). ggplot2: Elegant graphics for data analysis. *Journal of Statistical Software*, 35(1), 1–3. https://doi.org/10.18637/jss.v035.b01

Vinson, J., & Vinson, J. M. (1969). The Saurian fauna of the Mascarene Islands. *The Mauritius Institute Bulletin*, 6, 203–320.

Wang, Y., Naumann, U., Wright, S. T., & Warton, D. I. (2012). Mvabund—an R package for model-based analysis of multivariate abundance data. *Methods in Ecology and Evolution*, 3(3), 471–474. https://doi.org/10.1111/j.2041-210X.2012.00190.x

Zuël, N. (2009). *Ecology and Conservation of an Endangered Reptile Community on Round Island, Mauritius*. University of Zurich.

**SUPPORTING INFORMATION**

Additional supporting information may be found in the online version of the article at the publisher’s website.

**How to cite this article:** Tercel, M. P. T. G., Moorhouse-Gann, R. J., Cuff, J. P., Drake, L. E., Cole, N. C., Goder, M., Mootooocurpen, R., & Symondson, W. O. C. (2021). DNA metabarcoding reveals introduced species predominate in the diet of a threatened endemic omnivore, Telfair’s skink (*Leiolopisma telfairii*). *Ecology and Evolution*, 00, 1–14. https://doi.org/10.1002/ece3.8484