The inappropriate use of time-to-independence biases estimates of activity patterns of free-ranging mammals derived from camera traps

Christopher Peral | Marietjie Landman | Graham I. H. Kerley

Abstract
Measuring and comparing activity patterns provide key insights into the behavioral trade-offs that result in animal activity and their extrinsic and intrinsic drivers. Camera traps are a recently emerged source of data for sampling animal activity used to estimate activity patterns. However, nearly 70% of studies using such data to estimate activity patterns apply a time-to-independence data filter to discard appreciable periods of sampling effort. This treatment of activity as a discrete event emerged from the use of camera trap data to estimate animal abundances, but does not reflect the continuous nature of behavior, and may bias resulting estimates of activity patterns. We used a large, freely available camera trap dataset to test the effects of time to independence on the estimated activity of eight medium- to large-sized African mammals. We show that discarding data through the use of time-to-independence filters causes substantial losses in sample sizes and differences in the estimated activity of species. Activity patterns estimated for herbivore species were more affected by the application of time-to-independence data filters than carnivores, this extending to estimates of potential interactions (activity overlap) between herbivore species. We hypothesize that this pattern could reflect the typically more abundant, social, and patch-specific foraging patterns of herbivores and suggest that this effect may bias estimates of predator–prey interactions. Activity estimates of rare species, with less data available, may be particularly vulnerable to loss of data through the application of time-to-independence data filters. We conclude that the application of time-to-independence data filters in camera trap-based estimates of activity patterns is not valid and should not be used.

Keywords
behavior, camera traps, daily activity, overlap, pseudoreplication, Serengeti

Taxonomy Classification
Behavioural ecology
1 | INTRODUCTION

The daily activity patterns of an animal reflect its phylogeny and the risks and rewards of activity or inactivity that determine fitness (Halle, 2000; Roll et al., 2006). Measuring and comparing activity patterns provides key insights into the behavioral trade-offs that result in activity, such as food availability, mating opportunities, physiological processes, predation risk, and environmental constraints (Owen-Smith, 1998; Tambling et al., 2015; Weyer et al., 2020; Zaman et al., 2022). These insights, together with the emerging availability of abundant activity data from camera traps, have led to renewed interest in describing the free-ranging activity patterns of species and populations and comparing these between groups (e.g., predators and their prey, those at risk of predation vs. those not, and between time periods—Delisle et al., 2021; Diete et al., 2017; O’Connell et al., 2011; Rowcliffe et al., 2014; Smith et al., 2020; Zaman et al., 2022).

However, the trend in the literature is for such camera trap-based estimates of activity patterns to approach activity as a discontinuous rather than a continuous state. This occurs by separating activity data (captured images) into discrete events by applying a time-to-independence filter and discarding all the images of a particular individual (or species) within this time-to-independence interval for each camera. Nearly 70% of the open-access publications on free-ranging animal activity patterns that we reviewed (Web of Science: 90 of 134 open access publications between 1998 and 2021) apply such a time-to-independence filter to camera trap data (Figure 1, Table S1). These filters are usually arbitrary (i.e., lacking a rationale), although avoiding pseudoreplication may be invoked (e.g., Zaman et al., 2022). They typically are 30 min duration, but may extend to 60 min and even 24 h. This can lead to discarding activity data for appreciable portions of the 24-h cycle and will likely influence the ensuing estimates of activity patterns. This is analogous to the previously used approach of discarding autocorrelated location data in radio-tracking studies, a practice that introduces biases in animal home range estimates (de Solla et al., 1999).

This time-to-independence filter approach contrasts strongly with traditional estimates of activity patterns that record and express activity as continuous and use all records of activity to quantify activity. This is epitomized by Aschoff (1954), who defined animal activity as “an animal is active when it moves parts of its body or moves itself.” Altmann (1974), in the classic study on measuring behavior, would define activity as a “state” (i.e., the animal is either active or inactive), not an “event,” and catered for measuring this through focal animal sampling that yields a continuous record of behavioral states (and the occurrence of events). There is an extensive body of literature that analyzes the activity patterns of animals, using, for example, data from direct observations (Davies & Skinner, 1986), records of animals breaking infrared light beams or altering conductance in an arena (Perrin, 1981; Smit & Langman, 1974), wheel running (Siepka & Takahashi, 2005), or implanted accelerometers that record movement (Weyer et al., 2020). All these studies use records of activity at the highest resolution (i.e., shortest interval between records) possible, and none of them discard activity records from their analyses.

How did this disjunction between established approaches for quantifying activity patterns and the camera trap time-to-independence approach come about? The use of time-to-independence filtering appears to stem from its use in determining animal abundances and densities, where multiple images of the same individual cause inflated estimates of abundance and density (Green et al., 2020; O’Connell et al., 2011; Wearn & Glover-Kapfer, 2017). However, activity differs from the discrete nature of the occurrence of individuals or groups of animals. The absence of data to show that an animal is active during an observation period infers that it is inactive (the alternative state, following Altmann, 1974). Thus, the filtering and removal of activity data mean that the observer effectively decides that the animal is
inactive in this period, discarding meaningful information on animal activity. This may lead to biases in our estimation of the activity patterns of animals and therefore also in our ability to detect changes in activity in response to conspecifics, predators, or competitors, food availability, or physiological constraints. Currently, there is no conceptual or empirical information on the influence of the time-to-independence approach on estimates of animal behavior (or more specifically, animal activity patterns) from camera traps.

Using a large, freely available camera trap dataset from the Serengeti National Park, Tanzania (Swanson et al., 2015), we explored whether the use of time-to-independence filters alters estimates of the activity patterns of eight African mammal species and interpretations of the interactions between these species. We chose these species to represent a suite of traits (Table 1), including mode of activity (diurnal, nocturnal, crepuscular), social structure (solitary, gregarious), and trophic guild (carnivores, herbivores). Our approach was to estimate the activity patterns of each species using different intervals of time to independence and then to compare these patterns with species and estimate interactions (overlap) between species across different times to independence. We hypothesize that the use of time-to-independence filters will result in an underestimation of activity within species and of overlap in activity between species. This would be particularly relevant during peak activity periods when records of activity (images) are frequent and occur close together in time, hence applying time-to-independence filters would discard the most data and bias estimates downward.

2 | METHODS

2.1 | Dataset

To determine the influence of the interval of time to independence on mammal activity patterns, we used records of animal activity (images captured by motion-triggered camera traps) collected by the long-term Snapshot Serengeti project (Swanson et al., 2015, https://www.zooniverse.org/projects/zooniverse/snapshot-serengeti). The project comprises 225 motion-triggered camera traps placed in a 1125 km² grid in the center of Serengeti National Park, Tanzania. Cameras are active throughout the 24-h cycle, generally capturing bursts of images (three images per burst) within the first few seconds (1–10 s) of detected motion (Swanson et al., 2015). By May 2013, the project had produced ~1.2 million images, with the species identities classified by citizen scientists. We extracted data (comprising 164,509 images in total) on the eight mammal species collected between July 2010 and May 2013 for our study (Table 1).

We tested the influence of six time-to-independence intervals commonly used in the published literature: 1, 5, 10, 15, 30, and 60 min (Figure 1), with data from the longer intervals nested within the shorter intervals. Data for each interval were selected by sorting images by camera and species and removing images of the same species at a camera within the specified time interval. In all tests, we used the 1-min time-to-independence interval (rather than 0 min) as the base case (or control) for comparison. This was necessary due to possible differences in the number and duration of detection bursts (used to improve species identifications—Forrester et al., 2016) between camera models (Swanson et al., 2015).

2.2 | Data analysis

We used R 3.6.2 for all analyses (R Core Team, 2019) with a significance level of \( \alpha = 0.05 \) for statistical tests. To assess the adequacy of our sample sizes, we plotted hourly accumulation curves of activity for each mammal species and time-to-independence interval. We visually assessed the shapes of the accumulation curves, expecting them to stabilize once the relationship between activity (i.e., records in the hour recorded as active) and the cumulative number of images reached an asymptote. Accumulation curves were fitted using the R-library vegan (Oksanen et al., 2019). To determine where the accumulation curve reached an asymptote, we estimated breakpoints (two given the shape of the accumulation curve) with a segmented regression using the segmented library (Muggeo, 2010; Toms & Lesperance, 2003). Adequate sampling was achieved when the number of available images exceeded the number of images at the second breakpoint (i.e., at the asymptote).

To determine the daily activity patterns of each species for each interval of time to independence, we fitted non-parametric kernel density functions (Meredith & Ridout, 2014; Ridout & Linkie, 2009). To delimit broad activity peaks (i.e., where records of activity are high), we visually assessed the shape of the activity density curves. Two broad shapes emerged (i.e., activity peak around midday and activity peaks around dawn and dusk), which roughly separated between the trophic guilds. Thus, for ease of comparison, we delimited hours of peak activity for the herbivores between 11:00 and 13:00 and for the carnivores between 05:00 and 07:00 and again between 20:00 and 22:00 (i.e., reflecting their crepuscular mode of activity). Activity peaks were then described as the number of images recorded in these peak periods of activity and expressed as a percentage of the total number of images.

The coefficient of overlap (\( \Delta \)), implemented in the overlap library, was used to estimate the degree of similarity in daily activity (Meredith & Ridout, 2014; Ridout & Linkie, 2009) within and between species and with time to independence. The coefficient of overlap ranges from 0 to 1, where 0 indicates no overlap in activity and 1 indicates complete overlap. Confidence intervals (95%) for coefficients of overlap were calculated using at least 1000 bootstraps. Model parameters were set according to the recommendations of Ridout and Linkie (2009) and Meredith and Ridout (2014) throughout. We compared pairwise overlap in activity patterns statistically with a Watson \( U^2 \) test (circular library; Agostinelli & Lund, 2017; Zar, 2010).
RESULTS

3.1 Adequacy of sample sizes

In total, we extracted 164,509 images from the Snapshot Serengeti Project dataset. Available sample sizes varied widely across species. These declined with the application of increasing intervals of time to independence (Table 1). In all cases, accumulation curves reached an asymptote, and the total number of available images exceeded the number of images at the asymptotes (i.e., the second breakpoint of the segmented regression; Table S2). This confirmed adequate sampling to describe the activity patterns of all study species for each interval of time to independence.

### TABLE 1

| Study species | Social structure | Trophic guild | Total no. images | Time interval |
|---------------|-----------------|---------------|------------------|---------------|
|               |                 |               |                  | 1-min | 5-min | 10-min | 15-min | 30-min | 60-min |
| Wildebeest Connochaetes taurinus | Gregarious | Herbivore | 100,660 | 100,179 | 20,395 | 14,594 | 12,159 | 9224 | 7319 |
| Gazelle Gazella thomsoni | Gregarious | Herbivore | 41,420 | 41,349 | 19,367 | 15,647 | 13,932 | 11,451 | 9563 |
| Buffalo Syncerus caffer | Gregarious | Herbivore | 13,672 | 13,444 | 5336 | 4504 | 4136 | 3792 | 3521 |
| Eland Tragelaphus oryx | Gregarious | Herbivore | 2689 | 2687 | 1015 | 989 | 966 | 838 | 381 |
| Hyena Crocuta crocuta | Gregarious | Carnivore | 5303 | 5303 | 3601 | 3461 | 3379 | 2906 | 2486 |
| Serval Leptailurus serval | Solitary | Carnivore | 458 | 458 | 395 | 390 | 388 | 378 | 364 |
| Leopard Panthera pardus | Solitary | Carnivore | 228 | 228 | 184 | 181 | 180 | 178 | 175 |
| Caracal Caracal caracal | Solitary | Carnivore | 79 | 79 | 63 | 63 | 63 | 62 | 61 |

3.2 Within-species effects

Using time-to-independence filters predictably caused a loss of activity data across species (Table 1, Figure 2). However, the extent of the data loss varied between species: buffalo, gazelle, and wildebeest, species with large sample sizes, lost between 74% and 93% of their activity data when time to independence increased from 1 to 60 min. In contrast, species with smaller sample sizes, such as caracal, leopard, and serval, showed fewer data losses (between 21% and 23%) over the same time-to-independence intervals (Figure 2).

As expected, discarding activity data caused changes to the estimated daily activity patterns (Figure 3). In particular, increasing the interval of time to independence dampened the broad activity peaks.
(i.e., where records of activity are high) of the herbivores, but accentuated these for the carnivores (Table 2). Wildebeest, gazelle, and buffalo, for example, showed a decline in their midday activity peak, this by as much as 44% in the case of gazelle when time to independence increased from 1 to 60 min. In contrast, hyena and leopard gained more defined activity peaks at dawn (increasing by 9% and 17%, respectively) and dusk (increasing by 18% and 16%, respectively) over the same increments of time-to-independence intervals. For the herbivores, the incremental dampening of peak activity with time to independence reduced the degree of overlap between the respective activity curves (generated for each time interval) and the control (i.e., the 1-min interval) within each species (Table 3). That is, we detected a change in the overall activity pattern within each species with each time interval. In contrast, for the carnivores, the change in peak activity with time-to-independence filtering had little effect on overall activity patterns.

3.3 | Between-species effects

To explore whether using time-to-independence filters alters interpretations of the potential interactions between species, we overlaid the activity patterns of wildebeest × gazelle, wildebeest × eland, and buffalo × eland for each interval of time to independence. These species combinations were selected to demonstrate the range of potential effects of time to independence on interpretations of species interactions, as expressed by overlap in activity patterns. We included no carnivores here because we observed no striking changes in activity between our study species with changing time to independence (Table 3).

For the herbivores, activity overlap varied between the different species combinations and with time to independence (Figure 4). Activity overlap between wildebeest × gazelle increased steadily with an increase in time to independence, while activity overlap between wildebeest × eland appeared to be more resilient to the effects of the time interval. In contrast, activity overlap between buffalo × eland responded initially, but later appeared to stabilize with an increase in time to independence.

4 | DISCUSSION

Using a large, freely available camera trap dataset, our study explored the effects of time-to-independence data filters on the estimation of activity patterns in eight African large mammal species. We show that the application of these data filters alters our...
understanding of activity patterns of species and potential interactions between species and that these effects vary between species. The unusually large number of camera traps (225 traps) on which our study is based likely reduced the effect of discarding data in any time period (e.g., an animal’s activity peak), as near-contemporaneous activity data may have been collected for that species by another camera in the array. This suggests that studies using fewer camera traps (most such studies use fewer than 50) would be more vulnerable to the effects of discarding data through time-to-independence filters.

The application of time-to-independence filters causes the loss of activity data. In our study, herbivores lost more data than carnivores, and this was particularly striking during peak periods of activity (Figures 2 and 3, Tables 1 and 2). While the mechanism of this guild-level effect is not immediately clear, this likely reflects a combination of the effect of sample size and the distribution of samples over each 24-hr cycle, both of which varies with the species’ life history traits, social structure, and patch use. In the case of sociality, for example, these herbivores are typically more social and occur together in larger groups than the studied carnivores (Estes, 2012). This means that captured images on a camera can accumulate relatively quickly within specific time intervals during peak periods of herbivore activity, leading to large sample sizes and hence a substantial cumulative discarding of samples with the application of time-to-independence filters. This is supported by the fact that the most social of the carnivores sampled (the spotted hyena) had the greatest cumulative loss of images among members of the carnivore guild (Table 1).

Species that lost the least amount of data in our study (the carnivores) are not only typically rare (so fewer individuals are recorded at longer intervals), but also share a similar behavior in actively searching for prey across the landscape (Estes, 2012), which likely influences their capture rate on camera traps. This contrasts with the typically more social, abundant, and patch-specific foraging herbivores. Thus, the latter would be more likely to be represented in repeated captures on a camera trap operating in their foraging patch, while carnivores are more likely to move through a patch quickly and so accumulate fewer images within a specified time. This would exert an asymmetrical effect of time-to-independence filtering on these guilds, based on their life history characteristics. Thus, not only is it important to recognize that time-to-independence filtering alters our estimation of animal activity, but also which species or guilds of species are more or less at risk of the biases associated with time to independence. The ability to generalize our observation that estimates of some species’ activity patterns may be more vulnerable than others to the application of time-to-independence filters needs to be tested with data from additional species, and including rare, social herbivores (with small sample sizes) and carnivores with large sample sizes. Based on these guild-level differences in the effect of applying time to independence (differential loss of sample sizes and changes in estimated activity patterns), the use of time-to-independence filters in comparative activity studies of, for example, predators and prey, will lead to misleading outcomes.

These findings are also important when estimating the activity patterns of rare species, which are less frequently captured on camera traps (Lama et al., 2019). Although our results suggest that for these species, discarding data (through the application of a time-to-independence filter) will have a smaller effect on the proportion of available data, this may still lead to inadequate sample sizes for estimating activity patterns. This would be particularly important if discarding of activity events leads to the misclassification of activity patterns (e.g., diurnal vs. crepuscular), and hence a misunderstanding

### TABLE 2
Change proportion of images in the broad activity peaks estimated for each study species with time to independence, together with the percentage change in activity between the 1- and 60-min intervals of time to independence.

| Study species | Hours of peak activity | % Images in the activity peak | % Δ Activity peak (1-min:60-min) |
|---------------|------------------------|-------------------------------|---------------------------------|
| | 1-min | 5-min | 10-min | 15-min | 30-min | 60-min | |
| Wildebeest Connochaetes taurinus | 11:00–13:00 | 19.1 | 13.0 | 12.6 | 12.7 | 12.6 | 12.7 | -34 |
| Gazelle Gazella thomsoni | 11:00–13:00 | 26.2 | 18.1 | 16.6 | 15.7 | 14.9 | 14.6 | -44 |
| Buffalo Syncerus caffer | 11:00–13:00 | 15.9 | 12.5 | 11.5 | 10.8 | 10.3 | 9.4 | -41 |
| Eland Tragelaphus oryx | 11:00–13:00 | 19.1 | 14.0 | 14.3 | 14.5 | 13.7 | 14.4 | -25 |
| Hyena Crocuta crocuta | 05:00–07:00 | 9.5 | 10.4 | 9.9 | 9.9 | 10.2 | 10.0 | 6 |
| | 20:00–22:00 | 12.9 | 15.2 | 15.7 | 15.6 | 15.3 | 15.8 | 22 |
| Serval Leptailurus serval | 05:00–07:00 | 13.3 | 13.4 | 13.3 | 13.1 | 13.0 | 12.4 | -7 |
| | 20:00–22:00 | 15.1 | 14.9 | 15.1 | 15.2 | 15.3 | 15.7 | 4 |
| Leopard Panthera pardus | 05:00–07:00 | 14.9 | 17.4 | 17.7 | 17.8 | 18.0 | 17.7 | 19 |
| | 20:00–22:00 | 10.5 | 12.0 | 12.2 | 12.2 | 12.4 | 12.6 | 19 |
| Caracal Caracal caracal | 05:00–07:00 | 20.3 | 15.9 | 15.9 | 15.9 | 14.5 | 14.8 | -27 |
| | 20:00–22:00 | 12.7 | 15.9 | 15.9 | 15.9 | 16.1 | 14.8 | 17 |

Note: Negative values indicate a dampening of the activity peak with increasing time to independence, and positive values indicate that the peaks are accentuated.
| Time interval | Statistic | Wildebeest (95% CI) | Buffalo (95% CI) | Gazelle (95% CI) | Eland (95% CI) | Hyena (95% CI) | Serval (95% CI) | Leopard (95% CI) | Caracal (95% CI) |
|---------------|-----------|---------------------|------------------|------------------|----------------|----------------|-----------------|-----------------|-----------------|
| 5-min         | Overlap   | 0.85 (0.84–0.86)    | 0.92 (0.91–0.94) | 0.85 (0.84–0.86) | 0.90 (0.89–0.92) | 0.93 (0.92–1.00) | 0.89 (0.85–0.98) | 0.83 (0.77–0.98) |
|               | Watson's $U^2$ | 47.03            | 2.65             | 35.61            | 2.47            | 1.81            | 0.03            | 0.04            | 0.06            |
|               | $p$       | <.001              | <.001            | <.001            | <.001          | <.001          | >.10            | >.10            | >.10            |
| 10-min        | Overlap   | 0.84 (0.82–0.84)    | 0.90 (0.89–0.92) | 0.81 (0.81–0.82) | 0.85 (0.81–0.88) | 0.90 (0.88–0.92) | 0.93 (0.92–1.00) | 0.89 (0.85–0.97) | 0.83 (0.76–0.98) |
|               | Watson's $U^2$ | 43.97            | 4.01             | 46.25            | 2.33            | 2.05            | 0.03            | 0.04            | 0.06            |
|               | $p$       | <.001              | <.001            | <.001            | <.001          | <.001          | >.10            | >.10            | >.10            |
| 15-min        | Overlap   | 0.83 (0.82–0.83)    | 0.89 (0.88–0.91) | 0.80 (0.79–0.81) | 0.86 (0.82–0.88) | 0.90 (0.88–0.92) | 0.93 (0.92–1.00) | 0.89 (0.84–0.98) | 0.83 (0.77–0.98) |
|               | Watson's $U^2$ | 40.84            | 4.95             | 51.18            | 2.16            | 2.21            | 0.03            | 0.04            | 0.06            |
|               | $p$       | <.001              | <.001            | <.001            | <.001          | <.001          | >.10            | >.10            | >.10            |
| 30-min        | Overlap   | 0.82 (0.80–0.82)    | 0.88 (0.86–0.89) | 0.77 (0.76–0.78) | 0.84 (0.80–0.87) | 0.90 (0.88–0.92) | 0.93 (0.92–1.00) | 0.89 (0.85–0.97) | 0.83 (0.76–0.98) |
|               | Watson's $U^2$ | 36.29            | 5.93             | 57.32            | 2.41            | 1.78            | 0.03            | 0.04            | 0.06            |
|               | $p$       | <.001              | <.001            | <.001            | <.001          | <.001          | >.10            | >.10            | >.10            |
| 60-min        | Overlap   | 0.81 (0.80–0.82)    | 0.87 (0.85–0.88) | 0.75 (0.74–0.76) | 0.85 (0.81–0.87) | 0.90 (0.89–0.92) | 0.93 (0.91–1.00) | 0.89 (0.85–0.97) | 0.83 (0.76–0.98) |
|               | Watson's $U^2$ | 32.30            | 6.96             | 58.97            | 2.12            | 1.57            | 0.03            | 0.03            | 0.06            |
|               | $p$       | <.001              | <.001            | <.001            | <.001          | <.001          | >.10            | >.10            | >.10            |
of the species’ ecology, resource use patterns, and response to environmental pressures.

In many camera trap studies, “burst images” are used on the initial trigger of motion to capture additional images to improve species identifications (Wearn & Glover-Kapfer, 2017). These burst images are user-defined, not a true sample of animal activity, and artificially increase activity density. This means that data from these initial image bursts cannot be used to estimate activity and must be discarded from activity pattern analyses. Thus, for studies where these burst images are not specifically needed for individual identification, we suggest this setting should not be used, thus allowing for the recording of activity in a more continuous fashion.

The use of time-to-independence filters in studies that estimate animal activity patterns is therefore challenged under the principle that activity is a continuous state, and all records of activity of the same individual (or species) represent meaningful information. Thus, camera trap images of the same individual or group are not pseudoreplicates (sensu Hurlbert, 1984), but rather valid records of animal activity. The question arises as to whether the use of time to independence in published studies of activity patterns led to biases in the findings of these studies? A sample of such studies shows a variable approach (Figure 1) to the application of time to independence. Some studies use lengthy intervals of time to independence, ranging from 30 to 60 min (e.g., Farris et al., 2015; Foster et al., 2013; Santos et al., 2019; Tambling et al., 2015). This may have biased their results, but the nature of these effects is unknown. Studies contrasting prey and predator activity (e.g., Foster et al., 2013; Tambling et al., 2015; Smith et al., 2020; Zaman et al., 2022) that apply time-to-independence filters may need to be revisited. This is due to the demonstrated differences in the effects of time to independence between these two guilds. A similar effect is likely present among comparisons of activity patterns of rare and abundant species, in, for example, studies on competition (c.f. Santos et al., 2019).

Furthermore, the variable and arbitrary application of the time to independence across these studies means that the derived estimates of activity patterns cannot be meaningfully compared between studies.

5 | CONCLUSION

The use of time-to-independence filters in camera trap studies that describe the activity patterns of free-ranging animals is conceptually not justified, discards valuable data, and biases our understanding of the estimated activity patterns of animals. This may lead to incorrect inferences being drawn from such studies, although the extent and nature of these are currently unknown. The convention within the published literature of using lengthy periods of time to independence (often exceeding 30 min) for such studies is therefore challenged.

Studies estimating activity patterns from camera trap data should not discard activity data by applying time to independence. Alternatively, such studies should specifically test the effects of applying time to independence on estimated activity patterns and how these estimates may respond to factors that influence activity. This will lead to a more realistic description of the activity patterns of animals based on camera trap data, allow comparisons with activity patterns derived using other approaches, and generate greater confidence in our understanding of the factors that influence these activity patterns.

AUTHOR CONTRIBUTIONS

Christopher Peral: Conceptualization (equal); formal analysis (lead); methodology (equal); visualization (equal); writing – original draft (equal); writing – review and editing (equal). Marietjie Landman: Conceptualization (equal); formal analysis (supporting); methodology
Christopher Peral ORCID
snaps hot- seren geti.
The	dataset	of	this	study	is	available	on
DATA AVAILABILITY STATEMENT
ments	on	a	draft	manuscript.
ing	ta
Global	Impact	Award	from	Google,	and	by	a	grant	from	the
Alfred	P.	Sloan	Foundation.	We	thank	Peter	Apps	for	valuable	com
ments	on	a	draft	manuscript.

CONFLICT OF INTEREST
The authors have no conflicts of interest to declare.

DATA AVAILABILITY STATEMENT
The dataset of this study is available on https://lila.science/datasets/snapshot-serengeti.

ORCID
Christopher Peral https://orcid.org/0000-0003-4739-5452
Marietjie Landman https://orcid.org/0000-0002-5500-2121
Graham I. H. Kerley https://orcid.org/0000-0003-2702-5200

REFERENCES
Agostinelli, C., & Lund, U. (2017). R package ‘circular’: Circular statistics (version 0.4–93). https://r-forge.r-project.org/projects/circular/
Altman, J. (1974). Observational study of behavior - Sampling methods. Behaviour, 49, 227–267.
Aschoff, J.(1954). Timers of animal daily activity. Die Naturwissenschaften, 54, 49–56 [in German].
Davies, R. A. G., & Skinner, J. D. (1986). Temporal activity patterns of springbok Antidorcas marsupialis and merino sheep Ovis aries during a Karoo drought. Transactions of the Royal Society of South Africa, 46, 133–147.
De Solla, S. R., Bonduriansky, R., & Brooks, R. J. (1999). Eliminating autocorrelation reduces biological relevance of home range estimates. Journal of Animal Ecology, 68, 221–234.
Delisle, Z. J., Flaherty, E. A., Nobbe, M. R., Wzientek, C. M., & Swihart, R. K. (2021). Next-generation camera trapping: Systematic review of historic trends suggests keys to expanded research applications in ecology and conservation. Frontiers in Ecology and Evolution, 9, 617996.
Diete, R. L., Meek, P. D., Dickman, C. R., Lisle, A., & Leung, L. K. (2017). Diel activity patterns of northern Australian small mammals: Variation, fixity, and plasticity. Journal of Mammalogy, 98, 848–857.
Estes, R. D. (2012). The behavior guide to African mammals: Including hoofed mammals, carnivores, primates. University of California Press.
Farris, Z. J., Gerber, B. D., Karpaty, S., Murphy, A., Andrianjakarivelo, V., Ratelolahy, F., & Kelly, M. J. (2015). When carnivores roam: Temporal patterns and overlap among Madagascar’s native and exotic carnivores. Journal of Zoology, 296, 45–57.
Forrester, T., O’Brien, T., Fegraus, E., Jansen, P. A., Palmer, J., Kays, R., Ahumada, J., Stern, B., & McShea, W. (2016). An open standard for camera trap data. Biodiversity Data Journal, 4, e10197.
Foster, V. C., Sammento, P., Sollmann, R., Törres, N., Jácomo, A. T. A., Negrôs, N., Fonseca, C., & Silveira, L. (2013). Jaguar and puma activity patterns and predator-prey interactions in four Brazilian biomes. Biotropica, 45, 373–379.
Green, A. M., Chynoweth, M. W., & Sekercioglu, C. H. (2020). Spatially explicit capture-recapture through camera trapping: A review of benchmark analyses for wildlife density estimation. Frontiers in Ecology and Evolution, 8, e563477.
Halle, S. (2000). Ecological relevance of activity patterns. In M. M. Caldwell, G. Heldmaier, O. L. Lange, H. A. Mooney, E. D. Shulze, & U. Sommer (Eds.), Activity patterns in small mammals (1st ed., pp. 67–89). Springer.
Hurlbert, S. H. (1984). Pseudoreplication and the design of ecological field experiments. Ecological Monographs, 54, 187–211.
Lama, S. T., Ross, J. G., Bista, D., Sherpa, A. P., Regmi, G. R., Suwal, M. K., Sherpa, P., Weerman, J., Shrestha Lama, S., Thapa, M., Poudyal, L. P., Paterson, A. M., Weerman, J., Bista, D., Poudyal, L. P., Lama, S. T., Suwal, M. K., Sherpa, A. P., Regmi, G. R., ... Thapa, M. (2019). First photographic record of marbled cat Pardofelis marmorata Martin, 1837 (Mammalia, carnivora, Felidae) in Nepal. Nature Conservation, 32, 19–34.
Meredith, M., & Ridout, M. (2014). Overview of the overlap package. R project (version 0.34). https://cran.r-project.org/web/packages/overlap/index.html
Muğgeo, V. M. R. (2010). Segmented: An R package to fit regression models with broken-line relationships. https://cran.r-project.org/doc/Rnews/
O’Connell, A. F., Nichols, J. D., & Karanth, K. U. (2011). Camera traps in animal ecology: Methods and analyses. Springer.
Oksanen, A. J., Blanchet, F. G., Friendly, M., Kindt, R., Legendre, P., McGlinn, D., Minchin, P. R., Hara, R. B. O., Simpson, G. L., Solymos, P., Stevens, M. H. H., Szoecs, E., & Wagner, H. (2019). vegan: Community ecology package. R package (version 2.5–5). https://CRAN.R-project.org/package=vegan
Owen-Smith, N. (1998). How high ambient temperature affects the daily activity and foraging time of a subtropical ungulate, the greater kudu (Tragelaphus strepsiceros). Journal of Zoology, 246, 183–192.
Perrin, M. (1981). Notes on the activity patterns of 12 species of southern African rodents and a new design of activity monitor. South African Journal of Zoology, 16, 248–258.
R Core Team. (2019). R: A language and environment for statistical computing. R Foundation for Statistical Computing. https://www.R-project.org/
Ridout, M. S., & Linkie, M. (2009). Estimating overlap of daily activity patterns from camera trap data. Journal of Agricultural, Biological, and Environmental Statistics, 14, 322–337.
Roll, U., Dayan, T., & Kronfeld-Schor, N. (2006). On the role of phylogenetic and phylogeographic in determining activity patterns of rodents. Evolutionary Ecology, 20, 479–490.
Rowcliffe, J. M., Kays, R., Kranstauber, B., Carbone, C., & Jansen, P. A. (2014). Quantifying levels of animal activity using camera trap data. Methods in Ecology and Evolution, 5, 1170–1179.
Santos, F., Carbone, C., Wearn, O. R., Rowcliffe, J. M., Espinosa, S., Moreira, M. G., Ahumada, J. A., Gonçalves, A. L. S., Trevelin, L. C., Alvarez-Loayza, P., Spironello, W. R., Jansen, P. A., Juen, L., Peres, C. A., Lima, M. G. M., Ahumada, J. A., Gonçalves, A. L. S., Trevelin, L. C., Alvarez-Loayza, P., ... Peres, C. A. (2019). Prey availability and temporal partitioning modulate felid coexistence in neotropical forests. PLoS One, 14, e0213671.
Siepka, S. M., & Takahashi, J. S. (2005). Methods to record circadian rhythm wheel running activity in mice. Methods in Enzymology, 393, 230–239.
Smit, H., & Langman, V. A. (1974). A capacitive method of measuring circadian activity in laboratory animals. Medical & Biological Engineering, 12, 831–836.
Smith, J. A., Suraci, J. P., Hunter, J. S., Gaynor, K. M., Keller, C. B., Palmer, M. S., Atkins, J. L., Castaneda, L., Cherry, M. J., Garvey, P. M., Huebner, S. E., Morin, D. J., Teckentrup, L., Weterings, M. J. A., &
Beaudrot, L. (2020). Zooming in on mechanistic predator-prey ecology: Integrating camera traps with experimental methods to reveal the drivers of ecological interactions. *Journal of Animal Ecology, 89*, 1365–2656.

Swanson, A., Kosmala, M., Lintott, C., Simpson, R., Smith, A., & Packer, C. (2015). Snapshot Serengeti, high-frequency annotated camera trap images of 40 mammalian species in an African savanna. *Scientific Data, 2*, 1–14.

Tambling, C. J., Minnie, L., Meyer, J., Freeman, E. W., Santymire, R. M., Adendorff, J., & Kerley, G. I. H. (2015). Temporal shifts in activity of prey following large predator reintroductions. *Behavioral Ecology and Sociobiology, 69*, 1153–1161.

Toms, J. D., & Lesperance, M. L. (2003). Piecewise regression: A tool for identifying ecological thresholds. *Ecology, 84*, 2034–2041.

Wearn, O. R., & Glover-Kapfer, P. (2017). *Camera-trapping for conservation: A guide to best-practices*. WWF conservation technology series 1, Woking, UK.

Weyer, N. M., Fuller, A., Haw, A. J., Meyer, L. C. R., Mitchell, D., Picker, M., Rey, B., & Hetem, R. S. (2020). Increased diurnal activity is indicative of energy deficit in a nocturnal mammal, the aardvark. *Frontiers in Physiology, 1*, 637.

Zaman, M., Roberts, N. J., Zhu, M., Vitekere, K., Wang, M., & Jiang, G. (2022). Temporal activity patterns of North China leopards and their prey in response to moonlight and habitat factors. *Ecology and Evolution, 12*, e9032.

Zar, J. R. (2010). *Biostatistical analysis* (5th ed., p. 944). Prentice Hall.

**SUPPORTING INFORMATION**

Additional supporting information can be found online in the Supporting Information section at the end of this article.

---

**How to cite this article:** Peral, C., Landman, M., & Kerley, G. I. H. (2022). The inappropriate use of time-to-independence biases estimates of activity patterns of free-ranging mammals derived from camera traps. *Ecology and Evolution, 12*, e9408. [https://doi.org/10.1002/ece3.9408](https://doi.org/10.1002/ece3.9408)