Gas flaring activity and black carbon emissions in 2017
derived from the Sentinel-3A Sea and Land Surface
Temperature Radiometer

Alexandre Caseiro1,2, Berit Gehrke1,3, Gernot Rücker4, David Leimbach4, and Johannes W. Kaiser1,5

1Max Planck Institute for Chemistry, Mainz, Germany
2Institute for Advanced Sustainability Studies, Potsdam, Germany
3University Museum, University of Bergen, Bergen, Norway
4Zebris GbR, Munich, Germany
5Deutscher Wetterdienst, Offenbach, Germany

Correspondence: Alexandre Caseiro (alexandre.caseiro@mpic.de)

Received: 10 June 2019 – Discussion started: 25 July 2019
Revised: 4 July 2020 – Accepted: 18 July 2020 – Published: 10 September 2020

Abstract. Gas flares are a regionally and globally significant source of atmospheric pollutants. They can be
detected by satellite remote sensing. We calculate the global flared gas volume and black carbon emissions in
2017 by applying (1) a previously developed hot spot detection and characterisation algorithm to all observations
of the Sea and Land Surface Temperature Radiometer (SLSTR) instrument on board the Copernicus satellite
Sentinel-3A and (2) newly developed filters for identifying gas flares and corrections for calculating both flared
gas volumes (billion cubic metres, BCM) and black carbon (BC) emissions (g). The filter to discriminate gas
flares from other hot spots uses the observed hot spot characteristics in terms of temperature and persistence. A
regression function is used to correct for the variability of detection opportunities. A total of 6232 flaring sites
are identified worldwide. The best estimates of the annual flared gas volume and the BC emissions are 129 BCM
with a confidence interval of [35, 419 BCM] and 73 Gg with a confidence interval of [20, 239 Gg], respectively.
Comparison of our activity (i.e. BCM) results with those of the Visible Infrared Imaging Radiometer Suite
(VIIRS) Nightfire data set and SWIR-based calculations show general agreement but distinct differences in
several details. The calculation of black carbon emissions using our gas flaring data set with a newly developed
dynamic assignment of emission factors lie in the range of recently published black carbon inventories, albeit
towards the lower end. The data presented here can therefore be used e.g. in atmospheric dispersion simulations.
The advantage of using our algorithm with Sentinel-3 data lies in the previously demonstrated ability to detect
and quantify small flares, the long-term data availability from the Copernicus programme, and the increased
detection opportunity of global gas flare monitoring when used in conjunction with the VIIRS instruments.
The flaring activity and related black carbon emissions are available as “GFlaresS3” on the Emissions of atmospheric
Compounds and Compilation of Ancillary Data (ECCAD) website (https://doi.org/10.25326/19, Caseiro and
Kaiser, 2019).
1 Introduction

Industrial gas flaring (GF) occurs when flammable gas is disposed of by burning, most commonly done at the tip of a stack. This can either take place as a measure for pressure relief or to dispose of unwanted gas. In the upstream oil and gas industry (UOG, the oil and gas industry sector that includes searching, exploring, drilling, and operating the wells of crude oil and natural gas from underground or underwater sources and bringing the resource to the surface) in particular, GF occurs when the associated gas can not be sold easily and is not used on-site for energy generation (Gervet, 2007; Ismail and Umukoro, 2012; Emam, 2015).

Especially in insufficiently developed energy markets, companies seem to be spared from enough economic or political incentives to collect or convert the gas (Rahimpour and Jokar, 2012; Ojijiagwo et al., 2016). Improvements in flare gas recovery systems have been recommended for more closely monitored facilities (Zolfaghari et al., 2017; Papailias and Mavroidis, 2018).

GF negatively impacts the immediate surroundings (Akinola, 2018; Aujoulat and Lison, 2011; Nwoye et al., 2014), heat stress (Anomohanran, 2012; Anomohanran, 2012; Ajao and Anurigwo, 2002). GF also impacts the environment on a wider scale through the emission of pollutants and greenhouse gases like carbon dioxide (CO$_2$), carbon monoxide, black carbon (BC), nitrogen oxides, polycyclic aromatic hydrocarbons, volatile organic compounds, and acid rain precursors (Obioh et al., 1994; Uzomah and Sangodoyin, 2000; Nwaichi and Uzazobona, 2011; Onun et al., 2014; Li et al., 2016).

It has been estimated that between 2003 and 2012 GF produced on average 304 Tg CO$_2$ yearly, representing 0.6 % of the global anthropogenic emission of CO$_2$ equivalent (Olivier et al., 2014). Another source estimates the contribution as 400 Tg (Emam, 2015). The recovery of flare gas can play a relevant role in improving sustainability and meeting emissions targets (Abdulrahman et al., 2015; Comodi et al., 2016). Elvidge et al. (2018) estimated that a suppression of GF could contribute up to 2 % of the CO$_2$ eq, global nationally determined contributions (NDCs) under the United Nations Framework Convention on Climate Change Paris Agreement. Some countries could meet or even surpass their NDC (Yemen, Algeria, and Iraq), while others could meet between one-third and almost all of their NDC (Gabon, Venezuela, Iran, and Sudan).

The BC emissions from GF are of particular importance because BC is a known carcinogen (Heinrich et al., 1994) and a short-lived climate forcer (IPCC, 2013). BC strongly affects environments such as the Arctic regions by lowering the albedo of snow-covered surfaces (Flanner et al., 2007; Stohl et al., 2013; Bond et al., 2013; Huang et al., 2015; Elvidge et al., 2018), which has an impact on the Earth’s radiative balance (Doherty et al., 2010; Quinn et al., 2007; Hansen and Nazarenko, 2004) and also adds to the Arctic amplification phenomenon (Serreze and Barry, 2011). The GF contribution to the BC global emissions was estimated to amount to 270 and 210 Gg in 2005 and 2010, respectively (Klimont et al., 2017). Regionally, Stohl et al. (2013) and Cho et al. (2019) showed that GF contributes half of the near-surface BC concentration in the Arctic and explains a significant fraction of Arctic warming. Fossil fuel burning was also found to be the main source of BC deposited on snow (Qi and Wang, 2019).

GF is considered an important component of atmospheric dispersion simulations (Elvidge et al., 2018) and climate modelling (Huang and Fu, 2016), as the impact of GF emissions extends beyond immediate environmental concerns. However, information on the amount of natural gas being disposed of through stack burning and the accrued emissions is sparse. Reporting is often not enforced, inconsistent, or otherwise not reliable.

Satellite remote sensing has been utilised for regional and global identification and characterisation of GF (e.g. Elvidge et al., 1997; Casadio et al., 2012b; a; Anejionu et al., 2014; Faruolo et al., 2014; Chowdhury et al., 2014; Anejionu et al., 2015; Faruolo et al., 2018). The advantages and limitations of satellite remote sensing for the observation and monitoring of GF have recently been reviewed by Anejionu (2019). The most frequently cited system is NOAA’s VIIRS (National Ocean and Atmospheric Administration’s Visible Infrared Imaging Radiometer Suite) Nightfire (VNF) data set (see https://ngdc.noaa.gov/eog/viirs/download_viirs_fire.html, last access: 2 February 2020), developed by Elvidge et al. (2013, 2016) and still under active development (Elvidge et al., 2019). VNF, which succeeded a system based on the OLS (Operational Linescan System) instrument, which has been flying on board the DMSP (Defense Meteorological Satellite Program) satellites since the 1970s (Elvidge et al., 1997, 2001), provides a consistent global survey of flaring sites and GF volumes from 2012 onwards (https://ngdc.noaa.gov/eog/viirs/download_global_flare.html, last access: 2 February 2020).

Caseiro et al. (2018) describe an adaptation and extension of the VNF algorithm, with which observations of the Sea and Land Surface Temperature Radiometer instrument (SLSTR) on board the Copernicus Sentinel-3 satellites can be analysed as well. The algorithm detects and quantifies hot sources, including gas flares, using the night-time readings of the shortwave infrared (SWIR), mid-infrared (MIR), and thermal infrared (TIR) channels. SLSTR observations (night-time overpasses at 22:00 LST, local solar time) complement those of the VIIRS instruments (00:40 LST on JPSS-1/NOAA-20 and 01:30 LST on Suomi-NPP; our analysis pre-dates the launch of the latter, whose measurements are not included in the present work) by filling observation gaps in the time series. Both instruments are planned to allow long term data availability.

Here, we present a new data set of global GF sites, volumes (billion cubic metres, BCM), and BC emissions (g),
which is based on the observation by Sentinel-3A in 2017. Section 2 describes the newly developed methods for identifying gas flares among the observed hot sources, correcting for intermittent observations opportunities, and dynamically determining appropriate BC emission factors from the observations. To the best of our knowledge, it is the first time that the operational state of the flare, i.e. its temperature as determined from a satellite-based observation, is used to assign its BC emission factor. The results are presented in Sect. 3, including a comparison with other published methods (the VIIRS Nightfire product and the SWIR-radiance method), and the conclusions are summarised in Sect. 5.

2 Methods

2.1 Hot spot detection and characterisation

The steps involved in the study presented here are outlined in Fig. 1. We base our work on the persistent hot spot detection and characterisation algorithm described in Caseiro et al. (2018) to process a full year (2017) of Sentinel-3A’s Sea and Land Surface Temperature Radiometer (SLSTR) data. A hot event at temperatures typical of a gas flare (1200–2500 K) will produce a local maximum in the night-time readings of the shortwave and possibly of the mid-infrared (SWIR and MIR) channels of SLSTR. The SWIR band centred at 1.61 µm (S5) is closest to the expected spectral radiance maximum and serves as the primary detection band. Hot SWIR and MIR pixels are searched for using a contextual approach and their radiances are extracted. The radiometrically calibrated radiances of the different bands are used to fit the sum of two Planck curves: one representing the hot source and the other the cool background. Both curves are fitted to the night-time spectral infrared (shortwave, mid-wave and thermal infrared: SWIR, MIR, and TIR) observations to characterise temperature (T) and area (A) of the observed gas flares. Detections are divided into spurious signals (if the temperature retrieval is outside the 500–5000 K range) and valid hot spots. Hot spots are further divided into high accuracy (detection in at least another band besides S5, i.e. S6 or S7 or F1, meaning that with the LWIR there are at least 4 points for the Planck curve fit and at least 3 cloud-free pixels in the background to compute the background radiances) and low accuracy. Subsequently, the radiative power (RP) of the flare is calculated with the Stefan–Boltzmann equation.

While in principle the methodology used is based on VNF developed for Visible Infrared Imaging Radiometer Suite (VIIRS) (Elvidge et al., 2013, 2016), it differs by (1) analysing the radiances of clusters of contiguous hot pixels instead of treating spatial maxima as individual pixels and (2) using the TIR channels when fitting the sum of the two Planck curves. By considering all the hot pixels and therefore using all the information on the spatial scale, we expect our representation to be more realistic for the characterisation of large gas flare arrays. The use of the TIR channels implies a more stringent constraining of the background temperature. These options may lead to the retrieval of a lower flaring temperatures than those retrieved using the VNF approach (Caseiro et al., 2018). The lack of ground truth hinders the evaluation of the accuracy of the different options. Caseiro et al. (2018) already tested the method within a limited time span using oil- and/or gas-producing regions and compared the results to the VNF “flares only” product. The results showed good agreement of the hot source detection when investigating persistent hot spots with the advantage of the Sentinel-3A’s SLSTR algorithm in detecting and quantifying smaller flares, due to the night-time availability of a second SWIR channel. The characterisation in terms of temperature, area, and radiative power reached similar values. However, temperatures were slightly lower, while areas and radiative power were slightly larger in the SLSTR-based results. Note that VIIRS Nightfire also used two SWIR channels since January 2018 and that the detections are now conducted by two VIIRS instruments (at 00:40 and 01:30 LST).

2.2 Hot spot classification

2.2.1 Volcano filter

As discussed in (Caseiro et al., 2018), the criteria adopted to filter GFs are not enough to avoid false detections due to volcanoes, suggesting the use of masks. In order to filter out volcanic activity, we use the data available from the Global Volcanic Program of the Smithsonian institution (Venzke, 2013). A mask at 0.025° resolution was calculated from all eruptions in 2017. It also contains buffering zones stretching to 0.25° distance from the volcanic centre coordinates because these coordinates may not correspond to the location of the thermal radiation emission: many volcanoes do not consist of a single edifice, and many individual eruptive fissures through which lava erupts may be present in a volcanic field (Siebert et al., 2010).

2.2.2 Discrimination of gas flares from other industrial hot sources

In order to discriminate gas flares from other night-time hot spots we investigate persistence and retrieval temperature filtering.

We project all the hot spot detections onto a 0.025° × 0.025° global grid. This is a spatial resolution that reflects the reality on the ground, where several flares, separated by tens to hundreds of metres, may be operated within a single oil and gas facility, while facilities are tens of kilometres apart. This resolution is also adequate for the pixel footprint of medium-resolution imagers (Facchinelli et al., 2019). The following quantities are then computed for each grid cell: the number of hot spot detections (nObs); the number of high-accuracy (Caseiro et al., 2018) hot spot detections (nObsHA); the minimum, maximum, average, and standard deviation of the high-accuracy hot spot retrieved temperature (T), area
Figure 1. Flowchart of the methodology used in the present work. The starting point is the hot spot algorithm detection described in Caseiro et al. (2018) which, after applying a filter on the retrieved temperature ($500 \, K < T < 5000 \, K$), produces a database of valid hot spot detections. For each of these valid hot spot detections, the activity (billion cubic metres, BCM) and black carbon (BC) emissions are computed. The valid hot spot detections are further filtered: detection in at least another band besides S5 (S6 or S7 or F1, meaning that with the LWIR there are at least 4 points for the Planck curve fit, PCF) and at least 3 cloud-free pixels in the background to compute the background radiances. This step produces a database of high-accuracy hot spot detections. All hot spot detections are gridded into a global $0.025^\circ \times 0.025^\circ$ grid. For each grid cell, the following quantities are computed: the number of valid detections ($n_{\text{Obs}}$), high-accuracy detections ($n_{\text{ObsHA}}$), days of operation ($n_{\text{Ops}}$) (see below, Sect. 2.2.2), temperature, area, radiative power (minimum, average, maximum, and standard deviation), the date of the first and the last detection, and the sum of the activity (BCM) and BC emissions. A volcano filter is applied to the global grid in order to mask out volcanic activity (see Sect. 2.2.1). The last step is the discrimination of grid cells for flaring activity (see Sect. 2.2.2).

The temperature value used in the selection process of the discrimination strategy is based on Elvidge et al. (2016) and on the recent work by Liu et al. (2018), who derived gas flaring temperatures of 1000 to 2600 K from the VIIRS Nightfire database, depending on the type of operation (shale oil or gas and offshore, onshore, or refinery). According to these authors, most of the gas flares display temperatures between 1650 and 1850 K. However, temperatures can occasionally be as low as 1300 K. We therefore test for the minimum and/or for the maximum temperature for all the high-accuracy detections within a grid cell ($T_{\text{min}}$ and $T_{\text{max}}$, respectively). The temperature range reported by Elvidge et al. (2016) and Liu et al. (2018) (1650–1850 K) overlaps with particularly hot detections from the coal chemical industry and steel plants. Therefore, additional criteria are needed for identifying gas flares in the hot source data set, which are explored in the following paragraphs.

In order to select the discriminating strategy, we test several subsets of the gridded high-accuracy hot spot database. For each of the eight subsets described in Table 1, a sample of 100 random onshore grid cells complying to the defined thresholds have been tested by examining high-resolution...
Table 1. User’s accuracy (UA, %) and commission error (C, %) of the hot spot discrimination strategies considered. \( n_{\text{Obs}} \) is the number of hot spot detections within a grid cell, \( n_{\text{ObsHA}} \) is the number of high-accuracy hot spot detections within a grid cell, \( T_{\text{min}} \) is the minimum temperature retrieved among all the hot spots detected within a grid cell, \( T_{\text{max}} \) is the maximum temperature retrieved among all the hot spots detected within a grid cell, and \( n_{\text{cells}} \) is the number of grid cells that comply with the thresholds. We have tried eight combinations (discrimination strategies) of thresholds on those variables. Each column represents a tested discrimination strategy.

| Combination | no. 1 | no. 2 | no. 3 | no. 4 | no. 5 | no. 6 | no. 7 | no. 8 |
|-------------|-------|-------|-------|-------|-------|-------|-------|-------|
| \( n_{\text{Obs}} \geq \) | – | 2 | 3 | 4 | – | – | – | – |
| \( n_{\text{ObsHA}} \geq \) | – | 2 | 2 | 2 | 5 | 5 | 5 | 7 |
| \( T_{\text{min}} \, (K) \geq \) | 1000 | 1000 | – | – | – | – | – | – |
| \( T_{\text{max}} \, (K) \geq \) | 1400 | 1400 | 800 | 1200 | 1500 | 1800 | 1200 | 1500 |
| \( n_{\text{cells}} \) | 6733 | 5872 | 9469 | 6817 | 6232 | 5485 | 5527 | 5129 |
| UA | 84 ± 6 | 86 ± 8 | 60 ± 10 | 77 ± 13 | 85 ± 11 | 88 ± 10 | 73 ± 14 | 87 ± 11 |
| C | 7 ± 3 | 4 ± 2 | 19 ± 11 | 6 ± 4 | 3 ± 1 | 1 ± 1 | 8 ± 5 | 2 ± 1 |

imagery (Google Earth), and the locations are classified into four categories:

Flare (\( F \)): flaring site with visible flame;

Likely (\( L \)): industrial or oil extraction site with typical flaring infrastructure but no visible flame;

Unlikely (\( U \)): industrial site without typical flaring infrastructure;

No industry (\( N \)): e.g. agricultural or forested area.

The subsets are characterised in terms of user’s accuracy (UA, Eq. 1) and commission error (C, Eq. 2), where \( F, L, N, \) and \( U \) denote the number of grid cells in the corresponding category:

\[
\text{User’s accuracy} = \left( \frac{F + L}{2} \right) \pm \frac{L}{2}, \quad (1)
\]

\[
\text{Commission error} = \left( \frac{N + U}{2} \right) \pm \frac{U}{2}. \quad (2)
\]

The user’s accuracy is the accuracy from the point of view of the user. This metric represents the frequency with which a class on the map corresponds to the ground truth. In our case, UA lies between at least the verified flaring locations and at most the sum of the verified and the likely flaring locations. The commission error is calculated by reviewing the classified sites for incorrect classifications. In the present work, the commission error lies between at least the locations without any industry or infrastructure and at most the sum of the locations without any industry or infrastructure and the unlikely flaring locations.

In some industrial areas, facilities that use gas flares may be close to other hot spots, such as iron smelters or steel mills. In those cases, a commission error occurs when the flare is off and our methodology samples the other hot spot. The omission error can be divided into two categories: flares that were not detected as hot spots but which the discrimination strategy left out. The former will be the same for any discrimination procedure considered.

In order to discriminate gas flares from other hot spots, we discriminate hot spots based on their persistency (\( n_{\text{Obs}} \), number of hot spot detections, and \( n_{\text{ObsHA}} \), number of high-accuracy hot spot detections) and on their temperature time series (\( T_{\text{min}} \) and \( T_{\text{max}} \), the minimum and the maximum temperature retrieved among all the hot spots detected within a grid cell). We have tried eight combinations (discrimination strategies) of thresholds on those variables (Table 1). For each of the eight combinations, we examine high-resolution imagery for 100 random onshore locations (800 in total) in order to verify the presence of a gas flare. The goal is to maximise user’s accuracy (UA) and minimise commission error (C) while minimising the omission error (here, the variation in \( n_{\text{cells}} \) is used as a proxy). Discrimination strategy no. 5 (Table 1) was selected as the most suitable: detections located in grid cells where the yearly maximum temperature retrieval is above 1500 K and a persistency and quality criterion (\( n_{\text{ObsHA}} > 5 \)) is met are considered as originating from gas flares and are called flaring locations hereafter. This results in a relatively large number of grid cells with detections, i.e. low omission error, combined with a high user’s accuracy and a low commission error of 85 ± 11 % and 3 ± 1 %, according to Eqs. (1) and (2), respectively. Inserting a limitation on the minimum temperature or increasing the maximum temperature or the number of high-accuracy observations would not significantly increase the user’s accuracy or decrease the commission error, but would reduce the number of grid cells complying with the discrimination criteria by several hundred, which we interpret as an increase in the omission error. Although our discrimination strategy is satisfactory at the planetary level, global aggregating approaches may miss individual occurrences, in our case especially where the oil exploitation frontier is under expansion (Facchinelli et al., 2019).
 BCM assumes that the flare is constantly active: spot detections. The upper bound for the activity estimate assumes that the flare is only active on the $n$ maximum number of observations as a function of latitude. The latitude values are binned (10° bins; see Fig. 2). The function computes the maximum number of observations $n_{\text{Obs max}}$ per grid cell for a given latitude, which we assume expresses a continuous hot spot (365 d yr$^{-1}$). The number of days of operation $n_{\text{Ops}}$ is then estimated by scaling following Eq. (3). By this scaling, we approximately correct for the limitations of gas flaring detection from space (cloud cover and overpass frequency).

$$n_{\text{Ops}} = n_{\text{Obs}} \times \frac{365}{n_{\text{Obs max}}}$$

(3)

For the estimation of the flared volume, we applied the calibration derived by Elvidge et al. (2016) to each single detection within a flaring location grid cell. The calibration relationship uses a modified formulation of the radiative power to output the yearly flared volumes (in billion cubic metres, BCM), which is then scaled to a daily activity in m³.

The lower bound BCM$_{\text{min}}$ for the activity estimate assumes that the flare is only active on the $n_{\text{Obs}}$ days with hot spot detections. The upper bound for the activity estimate assumes that the flare is constantly active:

$$\text{BCM}_{\text{max}} = \text{BCM}_{\text{min}} \times \frac{365}{n_{\text{Obs}}}.$$  

(4)

The best activity estimate corresponds to assuming that the flare is active for $n_{\text{Ops}}$ days:

$$\text{BCM}_{\text{best}} = \text{BCM}_{\text{min}} \times \frac{n_{\text{Ops}}}{n_{\text{Obs}}}.$$  

(5)

The emissions of black carbon (BC) from gas flares are estimated using reported emission factors (EFs), using the traditional approach that the emission equals the multiplication of the activity (i.e. BCM) by the EF. Klimont et al. (2017) recognised the limited number of measurements of flaring emissions. Here, we maximise the use of the limited information available on the EFs by combining it with the individual observed flare characteristics.

EF data are scarce and the few published results span a considerable range of values; see McEwen and Johnson (2012), Stohl et al. (2013), Schwarz et al. (2015), Huang and Fu (2016), Weyant et al. (2016), and Klimont et al. (2017). In the present work, we apply the same concept of a varying EF as the one used by Klimont et al. (2017), who used the country of origin as an indication of the flare operation. However, instead of the physical location of the flare, we use the retrieved flaring temperature as an indication of the combustion completeness. Flaring temperatures close to the adiabatic flame temperature for natural gas (around 2500 K) are associated with more complete combustion and therefore lower BC emissions (0.5 g m$^{-3}$). On the other hand, low flaring temperatures (700 K and below) are associated with higher BC emissions (1.75 g m$^{-3}$). Between the two extremes, provided by the work of Klimont et al. (2017), the BC emission is scaled linearly as a function of the flaring temperature (see Fig. 3). To the best of our knowledge, this is the first time that operating conditions are taken into consideration when assigning the EF, and that the operating conditions are derived from direct observation.

The country-level BCM and BC estimates are computed by summing the individual flares estimates within the borders of each country and its exclusive economic zone.

With this methodology we estimate a wide range of possible activity (BCM) and BC emissions (g), where our best estimate falls between the “flaring only when there is a detection” (BCM$_{\text{min}}$ and BC$_{\text{min}}$) and “constant flaring” (BCM$_{\text{max}}$ and BC$_{\text{max}}$). We consider the range between BCM$_{\text{min}}$ and BCM$_{\text{max}}$ to be the confidence interval for the possible activity of a flaring site. Similarly, [BC$_{\text{min}}$, BC$_{\text{max}}$] is the confidence interval for the possible emission.

3 Results

3.1 Flaring locations

All the hot spots detected globally for 2017 using the algorithm as described in Fig. 1 are shown in Fig. 4. Their classification is shown in Fig. 5.

Noise, such as detections in the open ocean and the South Atlantic anomaly, is filtered out when considering only the high-accuracy hot spots as defined by Caseiro et al. (2018). There are 427 202 high-accuracy hot spots in the data set. Their distribution pattern suggests that they encompass a large range of hot sources, such as wildfires, volcanoes and industrial sources.
3.2 Flaring characteristics

Statistics on retrieved temperature are reported in Figs. 7 and 8.

The average temperature at the flaring locations approximately ranges from 950 to 2250 K. This is slightly smaller than the range 750–2500 K reported by Liu et al. (2018), who used VIIRS Nightfire data. This conforms to the findings of Caseiro et al. (2018). The distribution of average retrieved temperatures confirms the bi-modal distribution with modes around 1750 and 1200 K that has also been observed by VIIRS (Elvidge et al., 2013).

Statistics on retrieved radiative power are reported in Figs. 9 to 10.

The average radiative power at flaring locations range from a few tenths of a MW up to approximately 1 GW (Fig. 9), spanning the 5 orders of magnitude reported by Elvidge et al. (2013). Fisher and Wooster (2019) report a global FRP spanning from a few tens up to approximately 100 MW, with a median between 1 and 2 MW. Our results are very similar.

There were 197 610 detections at flaring locations in the SLSTR data set, most were clusters of up to 20 S5 pixels, as shown in Fig. 11.

3.3 Flared volumes

The activity for all the individual flaring locations is shown in Fig. 12. The most active flare burned 0.623 BCM in 2017 and is located south of Punta de Mata in Venezuela. This is also where Elvidge et al. (2016) found the most active flare for 2012, though more active (1.13 BCM).

Our best estimate for the global flared volume for 2017 is 129 BCM with a confidence interval of [35, 419] BCM.

Approximately half of the global flared volume originates from four countries only: Iraq (19 BCM), Iran (18 BCM), Russia (18 BCM), and Algeria (11 BCM) (Fig. 13).

3.4 BC emissions

The black carbon (BC) emissions for all the individual flaring locations are shown in Fig. 14. The most active flare emitted 0.35 Gg BC in 2017 (south of Punta de Mata).

Our best estimate for the global BC emissions from gas flaring is 73 Gg (lower estimate: 20 Gg; upper estimate: 239 Gg). As for the flared volume, four countries (see Fig. 15) account for more than half of that value: Iraq (11 Gg), Iran (11 Gg), Russia (10 Gg), and Algeria (6.0 Gg).

The global value determined here (73 Gg) is about one-third of the estimate from the Greenhouse Gas – Air Pollution Interactions and Synergies (GAINS) model (ECLIPSE V5a global emission fields) (Klimont et al., 2017): 270 and 210 Gg in 2005 and 2010, respectively. Such a decrease between 2010 and 2017 may be unlikely, given that the production within the UOG did not decrease in such a significant way. Rather than a decrease in the activity, or a shift in the emission factors (EF), the discrepancy might be due to how the EF is determined (see Sect. 2.3). Indeed, Klimont et al. (2017) used satellite-based retrievals for the flaring locations and their characteristics (VNF). Since the VNF methodology for the detection and characterisation of flares has several similarities with the one presented here, the differences in the computed emissions are more likely to come from a difference in the EF used. Therefore, we consider that part of the equation to be the most relevant in interpreting the difference between our results and the results from Klimont et al. (2017). Assigning the EF at the operational level for each detection seems more realistic than assigning it at the country level.

Our global estimate is larger than the global extrapolation made from a flaring emission study in the Bakken field by Weyant et al. (2016): 20 ± 6 Gg. The dynamic assignment
of the EF, based on the single temperature retrieval, should be closer to reality than globally extrapolating the EF measured at a single field. Huang and Fu (2016) used the VIIRS Nightfire data set for the flaring activity and emission factors derived from the flared gas chemical composition to compute the BC emissions. Their estimates range from approximately 140 to 200 Gg yr\(^{-1}\) between 1994 and 2012. VIIRS Nightfire was also the basis for the determination of atmospheric emissions from gas flaring by Doumbia et al. (2019) for the African continent. The authors computed a BC emission in
incident. Figure 17 compares the average retrieved flaring locations reported in the present work, 2964 are co-
set has at least one record. A flaring site is assigned to any grid cell for which the data 
retrieved flaring temperature > 1500 K and at least five high-
accuracy detections throughout the year. In the VNF data set, a flaring site is assigned to any grid cell for which the data set has at least one record.

Among the 10 185 VNF flaring locations and the 6232 flaring locations reported in the present work, 2964 are coincident. Figure 17 compares the average retrieved flaring temperature and the total BCM flared at these coincident flaring locations. Both methodologies retrieve temperatures roughly in the same range, with the points centred around a unique region and an observable trend in the comparison. Due to these characteristics and taking into consideration that a satellite-based observation is nothing but the snapshot of a moment in time, in this particular case the snapshot of an inconstant phenomenon, we find the correlation reasonably good. The SLSTR-based method retrieved temperatures that are more spread out over the values’ range than in the VNF product though. The BCM retrievals are well correlated with \( R^2 = 0.74 \). This shows that the differences, and presumably errors, in the temperature and flare area retrievals partly compensate for each other. The regression line shows that the BCM in the VNF data set is on average 16 % larger than the one retrieved by the SLSTR-based method.

A total of 7221 (70 %) VNF flaring locations and 3268 (52 %) SLSTR-based flaring locations are not coincident. The minimum distance from a VNF flaring location and the closest SLSTR-based flaring location, and vice-versa, are shown in Fig. 18.

We interpret the large difference in the number of detections as the effect of several factors: (1) small geolocation errors, (2) clustering versus local maxima detections, and (3) difference in the observation opportunities. In Table 2 we attempt to quantify the importance of these factors.

Distances of up to one grid cell (adjacent cells) may arise from small geolocation inaccuracies. This accounts for 2562 SLSTR-based locations and for 1507 VNF locations.

The Nightfire algorithm considers pixels that are local maxima and our methodology aggregates contiguous hot pixels within a unique cluster. These clusters may be very large, contain more than one local maximum and spread over grid cells, but only one cell is finally considered as its location. This was already indicated in Caseiro et al. (2018), where data retrieved from a natural gas and condensate production site in the Yamal peninsula (northern Siberia) were compared between VIIRS, SLSTR, and HSRS (Hot Spot Recognition System). In that particular case, the methodologies relying on clustering of hot pixels (SLSTR and HSRS) reported fewer gas flaring sites than VIIRS Nightfire. In the current database, about half of the 197 610 hot spots at flaring locations were clusters of up to 20 S5 pixels. At nadir, a S5 pixel has a ground footprint of 500×500 m². Therefore, 20 S5 pixels span a distance of up to 10 km, which corresponds to, at most, about four grid cells. Distances between VNF flaring locations that are not adjacent and less than four grid cells away from the closest SLSTR-based location may therefore arise from the clustering methodology. This is the case for 73 SLSTR-based locations and 1651 VNF locations.

The remaining 633 SLSTR-based locations and 4063 VNF locations were further apart. Of the distant locations in VNF not considered as flaring by our methodology (4063 grid cells), slightly more than half (2222) had at least one SLSTR detection, of which almost all (1964) had at least one detec-
Figure 7. Distribution of the average retrieved temperatures (K) for the flaring locations.

Figure 8. Average flaring temperature (K) at the 6232 flaring locations.

We thus conclude that, at VNF distant locations, a lack of enough detections and not the temperature hinders our methodology to classify those grid cells as flaring locations. Since the overpasses are not that far apart in time, cloudiness cannot account for this. The remaining factors affecting the detection opportunity are the overpass time and the swath. The difference in the overpass time between the sensors may play some role: the Suomi-NPP platform observes later during the night, thus possibly having more opportunities to detect gas flares at night. However, the effect of the different swaths of the instruments is likely more important.
Table 2. Comparison between the number (n) of flaring locations by VNF and the SLSTR-based method: total number, coincident, adjacent (< 2 grid cells away), clustering adjacent (≥ 2 and ≤ 4 grid cells away), and distant (≤ 4 grid cells away) locations. The sum of the activity (in BCM) for each is also shown.

| Distance d | Grid cells | Total flaring locations | Coincident locations | Adjacent locations | Clustering adjacent locations | Distant locations |
|------------|------------|-------------------------|----------------------|--------------------|--------------------------------|------------------|
| SLSTR-based | n          | 6232                    | 2964                 | 2562               | 73                            | 633              |
|            | BCM        | 129                     | 78                   | 39                 | 0.9                           | 10               |
| VNF        | n          | 10185                   | 2964                 | 1507               | 1651                          | 4063             |
|            | BCM        | 151                     | 132                  | 7.3                | 3.1                           | 8.7              |

VIIRS has a much larger swath (3040 km against 1420 km for SLSTR) and has therefore more opportunities for detection. The detection opportunity is more important if the flare exhibits little activity and, indeed, the 4063 VNF distant flaring locations account for just 8.7 BCM in the VNF inventory. The methodology presented here could be adjusted and the threshold in the number of high-accuracy observations lowered, but this would come at the cost of increasing false positives (see Sect. 2).

The very large majority of the VNF-based activity was detected at locations coincident with SLSTR-based locations (87 % at coincident locations and 92 % at coincident or adjacent locations). A total of 98 % of the SLSTR-based activity was detected at coincident or adjacent locations. This shows that both methodologies detect similar global levels of activity (151 BCM computed by VNF and 129 BCM for the present work) and this at the same locations, despite the large discrepancy in the number of flaring locations.

Table 2 shows that the difference in the computed global activity comes mostly from coincident and adjacent flaring locations, which indicates that the computation of the activity levels is more likely the source of the discrepancy (22 BCM globally) rather than the detection itself. One possibility is the scaling we use to obtain the annual activity from the individual detections (see Sect. 2, Eq. 3, and Fig. 2). Another source of discrepancy could be the lower temperature retrieved by our method, which considers the TIR for the Planck curve fitting.

Figure 21 shows the global distributions of the average retrieved flaring temperatures. The generally lower temperature in the present work can be traced back to the clustering methodology we take and was discussed in our previous paper (Caseiro et al., 2018). The use of the TIR channels in the dual Planck curve fitting may also lower the retrieved hot source temperature.

3.6 Comparison with the SWIR-radiance method

Fisher and Wooster (2018) developed a methodology to derive the radiative power from flares using a single SWIR channel, analogous to a previous methodology which used the MWIR radiance to derive the fire radiative power in the case of landscape fires (Wooster et al., 2005). The methodology is based on the near-constant ratio between the SWIR spectral radiance and the total emitted radiance at typical flaring temperatures (i.e. 1600–2200 K). The radiative power can be computed, using a sensor-specific optimised relationship and without the input of the emitter’s temperature, with a theoretical precision of 13.6 %.

Fisher and Wooster (2019) used the SWIR-radiance methodology to derive a global time series of flaring activity using (A)ATSR and SLSTR data. For SLSTR, the data used spanned from May 2017 to May 2018. The identification of night-time thermal hotspots was conducted using a simple static threshold: pixels having a radiance above the instrumental noise level were considered as hot spots. The thermal hotspots are subsequently gridded onto a 1 arcmin global geographic grid (approximately 2 km). For SLSTR, a grid cell is considered a flaring location if there were at least 2 different months with hotspot observations. The authors re-
Figure 10. Average radiative power (MW) at the 6232 flaring locations.

Figure 11. Distribution of the number of S5 pixels per cluster for the 197,610 valid detections at the 6232 flaring locations. Note the log scale on the y axis.

port 10,428 locations, with a global distribution very similar to what we report here.

In order to compare the methodologies, we apply the Single-SWIR method (Fisher and Wooster, 2019) to our detections and compare the output (RP) to our results (197,610 valid detections at flaring locations). Since the BCM in both methodologies is derived from linear models, the differences seen in the RP comparison would be reproduced in a BCM comparison. We limit the comparison to detections for which the retrieved temperature falls in the range 1600–2200 and which were limited to one S5 pixel in order to comply with the pre-requisites of the SWIR-radiance approach. The SWIR-radiance approach also allows us to compute the activity. However, since the BCM in both methodologies is de-
derived from linear models, the differences seen in the RP comparison would be reproduced in a BCM comparison.

Figure 22 shows the comparison between the RP derived from the SWIR-radiance method and the Planck curve fitting followed by the application of the Stefan–Boltzmann equation. There is a good correlation, although the SWIR-radiance methodology outputs radiative power values about 25% lower than those from this work. The trend is more recognisable at lower temperatures, which might indicate that the relationship between the SWIR spectral radiance and the total emitted radiance deviates more from linearity at lower temperatures.

4 Data availability

The flaring activity and the related black carbon emission product are available with 0.025° × 0.025° resolution on the Emissions of atmospheric Compounds and Compilation of Ancillary Data (ECCAD, https://eccad3.sedoo.fr/#GFFlaringS3, last access: 1 June 2020; https://doi.org/10.25326/19, Caseiro and Kaiser, 2019) website (https://eccad3.sedoo.fr, last access: 1 June 2020) for use in, e.g. atmospheric composition modelling studies.

5 Conclusions

In this work we apply a previously described hot spot detection and characterisation procedure (Caseiro et al., 2018) to the 2017 observations of Copernicus Sentinel-3A’s Sea and Land Surface Temperature Radiometer (SLSTR) instrument. In our previous paper, only persistency was used to discriminate gas flares from other sources of heat emissions. Here, we refine the discrimination by updating the persistency criterion and introducing a temperature criterion, two characteristics of gas flares. The development of the refined discrimination strategy is reported. Validation through referencing with high-resolution images yields a detection accuracy of 85 ± 11 %, with a commission error of 3 ± 1 %.

Our methodology detects 6232 flaring sites worldwide in 2017. Over half of these are located in five countries only: Russia, the United States, Iran, China, and Algeria.

Additionally, we calculate the volume of flared gas based on a previously observed and published linear relationship between the flared volume and observed flare radiative energy. Subsequently, estimates of the associated black carbon (BC) emissions are calculated with a new dynamic emission factor parameterisation.
Figure 16. VIIRS Nightfire flaring locations for 2017. We considered every record in the VIIRS Nightfire data as a flare, i.e. “upstream” (10 009), “downstream oil” (718), and “downstream gas” (101). After gridding, 10 185 cells were populated and are considered flaring sites. A detailed spatial analysis between these locations and those found by our methodology is shown in Fig. 18.

Figure 17. Comparison of the average temperature and the total BCM retrieved by VIIRS Nightfire and the SLSTR-based method at the 2964 coincident flaring locations. The red line represents the 1:1 relationship. The dashed line represents the linear regression between both methodologies.

The best estimate of flared gas at the detected flaring sites is calculated by applying a latitude-dependent scaling factor that roughly describes the variation in observation opportunities due to cloud cover and overpass frequency. The confidence interval is specified with the following bounds: the upper bound is computed as if the flare would constantly burn for the whole year. The lower bound is computed as if the flare would burn on days with hot spot detections only. The spread of the confidence intervals and the accuracy of the best estimates could certainly be reduced in the future with a more sophisticated correction for different observation opportunities that accounts for the actual number of cloud-free and solar-uncontaminated overpasses of each individual flaring site.

Our analysis yields a result of 129 billion cubic metres (BCM) of gas flared in 2017 (best estimate) with a confidence interval of [35,419] BCM. Only four countries are responsible for half of the global gas flaring activity: Iraq, Iran, Russia, and Algeria. The most active flaring location is located in Venezuela and burned 0.62 BCM in 2017.

A comparison with the VIIRS Nightfire data set shows that the number of flaring sites assessed here is significantly lower (6232 vs. 10185). Three main reasons can lead to the discrepancy: (1) small geolocation errors may result in flares being
assigned to adjacent grid cells. (2) The clustering methodology employed for SLSTR may result in an assignment of gas flare arrays spanning several grid cells to only one central grid cell. (3) The relatively stringent persistency criterion of the SLSTR method certainly excludes some infrequently operation gas flares. The former two reasons result in a mismatch between the SLSTR-based and VNF locations of gas flares but the aggregated budgets by country are not affected. The stringency of the persistency criterion might affect 4063 of the 10 185 VNF locations and these have a relatively little contribution of 8.7 BCM to the annual total of 151 BCM in VNF. The incorporation of Sentinel 3B in our system would lead to more observation opportunities. The inclusion of this feature and an analysis for the following years is pending and we expect it to provide substantially more coincidences between the two data sets. The complementarity and consistency of both products indicate that a merged VIIRS/SLSTR product would provide substantial improvement of the global flaring activity and emissions.

We also compare our activity estimates to the SWIR-radiance method. Both methodologies show a good agreement for high-temperature detections, but a constant under-
estimation by the SWIR-radiance method for lower temperature detections.

We further quantify the black carbon (BC) emissions due to gas flaring. We assume that temperature is an indication of the completeness of combustion and use the retrieved flame temperature to determine the emission factor for each single detection, bounded by the range of previously published emission factors. We assume that a temperature close to the adiabatic flame temperature corresponds to the lower bound of the emission factor range considered, while the lowest temperature corresponds to the upper bound. The minimum, maximum, and best estimates for the BC emissions are computed in the same way as the flared volume. Our resulting best estimate for the emission of BC to the atmosphere by gas flaring in 2017 is 73 Gg with a confidence interval of [20, 239] Gg, with Iraq, Iran, Russia, and Algeria being responsible for roughly one-half of those emissions. The most active flaring location, which is also the one with the highest emissions, is estimated to have yearly emissions of 0.35 Gg of BC in 2017.

This study shows that the SLSTR instruments on-board the Copernicus Sentinel-3 satellites are well suited for the quantitative characterisation of the gas flares in terms of flame temperature, size, and radiative power, as well as BC emissions. This will allow us to increase the detection opportunity for gas flares in an observation system that comprises SLSTR and VIIRS instruments.

Author contributions. This work was conceptualized by AC and JWK. The methodology was developed by AC, JWK, and GR. The software was developed and written by AC and DL. Validation was conducted by AC, JWK, and BG. Formal analysis, investigation, and resources were performed by AC and BG. Data were curated by AC, BG, DL, GR, and JWK. The original draft was prepared and written by AC. The paper was reviewed and edited by AC, BG, JWK, and GR. Visualization was performed by AC. Supervision, project administration, funding acquisition were conducted by JWK and GR.

Competing interests. The authors declare that they have no conflict of interest.

Acknowledgements. The data were produced in the project "GFS3 – Identification of gas flares and quantification of their emissions using Sentinel-3 SLSTR" funded by the German Federal Ministry for Economic Affairs and Energy (BMWi, Bundesministerium für Wirtschaft und Energie) under contract no. FKZ 50EE1339. David Leimbach and Gernot Rücker also acknowledge support by BMWi through the ZIM project FireSense (FKZ 16KN052420). We acknowledge Daniel Fisher and Chris Elvidge for their comments. We acknowledge Mariapia Faruolo and the remaining two anonymous referees for their reviews.
Financial support. This research has been supported by the German Federal Ministry for Economic Affairs and Energy (BMWi, Bundesministerium für Wirtschaft und Energie, grant nos. FKZ 50EE1339 and FKZ 16KN052420).

Review statement. This paper was edited by Vinayak Sinha and reviewed by Mariapia Faruolo and one anonymous referee.

References

Abdulrahman, A. O., Huisingh, D., and Hafkamp, W.: Sustaintability improvements in Egypt’s oil & gas industry by implementation of flame gas recovery, J. Clean. Prod., 98, 116–122, https://doi.org/10.1016/j.jclepro.2014.11.086, 2015.

Ajao, E. A. and Anurigwo, S.: Land-Based Sources of Pollution in the Niger Delta, Nigeria, Ambio, 13, 442–445, 2002.

Akionu, O. A.: Resource Misgovernance and the Contradictions of Gas Flaring in Nigeria: A Theoretical Conversation, J. Asian Afr. Stud., 53, 749–763, https://doi.org/10.1177/0021909617722374, 2018.

Anejionu, O. C. D.: Rationale, historical developments and advances in remote sensing of gas flares, Int. J. Remote Sens., 40, 1–20, https://doi.org/10.1080/01431161.2019.1500880, 2019.

Anejionu, O. C. D., Blackburn, G. A., and Whyatt, J. D.: Satellite survey of gas flares: development and application of a Landsat-based technique in the Niger Delta, Int. J. Remote Sens., 35, 1900–1925, https://doi.org/10.1080/01431161.2013.879351, 2014.

Anejionu, O. C. D., Blackburn, G. A., and Whyatt, J. D.: Detecting gas flares and estimating flaring volumes at individual flow stations using MODIS data, Remote Sens. Environ., 158, 81–94, https://doi.org/10.1016/j.rse.2014.11.018, 2015.

Anomohanran, O.: Determination of greenhouse gas emission resulting from gas flaring activities in Nigeria, Energ. Pol., 45, 666–670, https://doi.org/10.1016/j.enpol.2012.03.018, 2012.

Bond, T. C., Doherty, S. J., Fahey, D. W., Forster, P. M., Berntsen, T., DeAngelis, B. J., Flanner, M. G., Ghan, S., Kärcher, B., Koch, D., Kinne, S., Kondo, Y., Quinn, P. K., Sarofim, M. C., Schultz, M. G., Schulz, M., Venkataraman, C., Zhang, H., Zhang, S., Bellouin, N., Guttikunda, S. K., Hopke, P. K., Jacobson, M. Z., Kaiser, J. W., Klimont, Z., Lohmann, U., Schwarz, J. P., Shindell, D., Storelvmo, T., Warren, S. G., and Zender, C. S.:Bounding the role of black carbon in the climate system: A scientific assessment, J. Geophys. Res.-Atmos., 118, 5380–5552, https://doi.org/10.1002/jgrd.50171, 2013.

Casadio, S., Arino, O., and Minchella, A.: Use of ATSR and SAR measurements for the monitoring and characterisation of night-time gas flaring from offshore platforms: The North Sea test case, Remote Sens. Environ., 123, 175–186, https://doi.org/10.1016/j.rse.2012.03.021, 2012a.

Casadio, S., Arino, O., and Serpe, D.: Gas flaring monitoring from space using the ATSR instrument series, Remote Sens. Environ., 116, 239–249, https://doi.org/10.1016/j.rse.2010.11.022, 2012b.

Caseiro, A. and Kaiser, J. W.: Global Gas Flaring activity and Black Carbon emissions, ECCAD, https://doi.org/10.25326/19, 2019.

Caseiro, A., Rücker, G., Tiemann, J., Leimbach, D., Lorenz, E., Frauenberger, O., and Kaiser, J. W.: Persistent Hot Spot Detec-
tion and Characterisation Using SLSTR, Remote Sensing, 10, 1118, https://doi.org/10.3390/rs10071118, 2018.

Cho, M.-H., Park, R. J., Yoon, J., Choi, Y., Jeong, J. I., Labzovskii, L., Fu, J. S., Huang, K., Jeong, S.-J., and Kim, B.-M.: A missing component of Arctic warming: black carbon from gas flaring, Environ. Res. Lett., 14, 094011, https://doi.org/10.1088/1748-9326/ab374d, 2019.

Chowdhury, S., Shipman, T., Chao, D., Elvidge, C., Zhizhin, M., and Hsu, F.: Daytime gas flare detection using Landsat-8 multispectral data, in: 2014 IEEE Geoscience and Remote Sensing Symposium, 13–18 July 2014, Quebec City, QC, Canada, 258–261, https://doi.org/10.1109/IGARSS.2014.6946406, 2014.

Comodi, G., Renzi, R. M., and Rossib, M.: Energy efficiency improvement in oil refineries through flare gas recovery technique to meet the emission trading targets, Energy, 109, 1–12, https://doi.org/10.1016/j.energy.2016.04.080, 2016.

Doherty, S. J., Warren, S. G., Grenfell, T. C., Clarke, A. D., and Brandt, R. E.: Light-absorbing impurities in Arctic snow, Atmos. Chem. Phys., 10, 11647–11680, https://doi.org/10.5194/acp-10-11647-2010, 2010.

Douvbia, E. H. T., Lioussse, C., Keita, S., Granier, L., Granier, C., Elvidge, C. D., Elguindi, N., and Law, K.: Flaring emissions in Africa: Distribution, evolution and comparison with current inventories, Atmos. Environ., 199, 423–434, https://doi.org/10.1016/j.atmosenv.2018.11.006, 2019.

Elvidge, C., Baugh, K., Hobson, V., Kühn, E., Krochl, Herbert Davis, E., and Cocero, D.: Satellite inventory of human settlements using nocturnal radiation emissions: a contribution for the global toolchest, Glob. Change Biol., 3, 387–395, https://doi.org/10.1046/j.1365-2486.1997.00115.x, 1997.

Elvidge, C. D., Imhoff, M. L., Baugh, K. E., Hobson, V. R., Nelson, I., Safran, J., Dietz, J. B., and Tuttle, B. T.: Night-time lights of the world: 1994–1995, ISPRS Journal of Photogramm., 56, 81–99, https://doi.org/10.1016/S0924-2716(01)00040-5, 2001.

Elvidge, C. D., Zhizhin, M., Hsu, F.-C., and Baugh, K. E.: VIIRS Nightfire: Satellite Pyrometry at Night, Remote Sensing, 5, 4423–4449, https://doi.org/10.3390/rs5094423, 2013.

Elvidge, C. D., Zhizhin, M., Baugh, K., Hsu, F.-C., and Ghosh, T.: Methods for Global Survey of Natural Gas Flaring from Visible Infrared Imaging Radiometer Suite Data, Energies, 9, 14, https://doi.org/10.3390/en9010014, 2016.

Elvidge, C. D., Bazilian, D. M., Zhizhin, M., Ghosh, T., Baugh, K., and Hsu, F.-C.: The potential role of natural gas flaring in meeting greenhouse gas mitigation targets, Energy Strateg. Rev., 20, 156–162, https://doi.org/10.1016/j.esr.2017.12.012, 2018.

Elvidge, C. D., Zhizhin, M., Baugh, K., Hsu, F.-C., and Ghosh, T.: Extending Nighttime Combustion Source Detection Limits with Short Wavelength VIIRS Data, Remote Sensing, 11, 395, https://doi.org/10.3390/rs11040395, 2019.

Emam, E. A.: Gas Flaring in Industry: an overview, Petroleum and Coal, 57, 532–555, available at: http://large.stanford.edu/courses/2016/ph240/miller1/docs/emam.pdf (last access: 1 May 2019), 2015.

Evangelou, N., Shevchenko, V. P., Yttri, K. E., Eckhardt, S., Sol-lum, E., Pokrovsky, O. S., Kokelev, V. O., Korobov, V. B., Lobanov, A. A., Starodumova, D. P., Vorobiev, S. N., Thompson, R. L., and Stohl, A.: Origin of elemental carbon in snow from western Siberia and northwestern Europe during
winter—spring 2014, 2015 and 2016, Atmos. Chem. Phys., 18, 963–977, https://doi.org/10.5194/acp-18-963-2018, 2018.

Facchinelli, F., Pappalardo, S. E., Codato, D., Dianтини, A., Della Fera, G., Crescini, E., and De Marchi, M.: Unburnable and Unleakable Carbon in Western Amazon: Using VIIRS Nightfire Data to Map Gas Flaring and Policy Compliance in the Yasuni Biosphere Reserve, Sustainability, 12, 58, https://doi.org/10.3390/su12100158, 2019.

Faruolo, M., Coviello, I., Filizzola, C., Lacava, T., Pergola, N., and Tramutoli, V.: A satellite-based analysis of the Val d’Agri Oil Center (southern Italy) gas flaring emissions, Nat. Hazards Earth Syst. Sci., 14, 2783–2793, https://doi.org/10.5194/nhess-14-2783-2014, 2014.

Faruolo, M., Lacava, T., Pergola, N., and Tramutoli, V.: On the Potential of the RST-FLARE Algorithm for Gas Flaring Characterization from Space, Sensors, 18, 2466, https://doi.org/10.3390/s18082466, 2018.

Fisher, D. and Wooster, M. J.: Shortwave IR Adaption of the Mid-Infrared Radiance Method of Fire Radiative Power (FRP) Retrieval for Assessing Industrial Gas Flaring Output, Remote Sensing, 10, S305, https://doi.org/10.3390/rs10020305, 2018.

Fisher, D. and Wooster, M. J.: Multi-decade global gas flaring change inventoried using the ATSR-1, ATSR-2, AATSR and SLSTR data records, Remote Sens. Environ., 232, 111298, https://doi.org/10.1016/j.rse.2019.111298, 2019.

Flanner, M. G., Zender, C. S., Randerson, J. T., and Rasch, P. J.: Present-day climate forcing and response from black carbon in snow, J. Geophys. Res.-Atmos., 112, D11202, https://doi.org/10.1029/2006JD008003, 2007.

Gervet, B.: Gas Flaring Emission Contributions to Global Warming, Master’s thesis, INSA, Lyon, France and Luleå University of Technology, Luleå, Sweden, available at: http://www.ltu.se/cms_fs/1.5035!/gasflaringreport-final.pdf (last access: 1 May 2019), 2007.

Hansen, J. and Nazarenko, L.: Soot climate forcing via snow and ice albedo, P. Natl. Acad. Sci. USA, 101, 423–428, https://doi.org/10.1073/pnas.0232715100, 2004.

Heinrich, U., Dungworth, D. L., Pott, F., Peters, L., Dasenbrock, C., Levens, K., Koch, W., Creutenberg, O., and Schulte, A.: The Carcinogenic Effects of Carbon Black Particles and Tar – Pitch Condensation Aerosol after Inhalation Exposure of Rats, Ann. Occup. Hyg., 38, 351–356, https://doi.org/10.1093/annhyg/38.inhaled_particles_VII.351, 1994.

Huang, K. and Fu, J. S.: A global gas flaring black carbon emission rate dataset from 1994 to 2012, Scientific Data, 3, 160104, https://doi.org/10.1038/sdata.2016.104, 2016.

Huang, K., Fu, J. S., Prihodko, V. V., Storey, J. M., Romanov, A., Hodson, E. L., Cresko, J., Morozova, I., Ignatieva, Y., and Cabaniss, J.: Russian anthropogenic black carbon: Emission reconstruction and Arctic black carbon simulation, J. Geophys. Res.-Atmos., 120, 11306–11333, https://doi.org/10.1002/2015JD023358, 2015.

IPCC: Climate Change 2013: The Physical Science Basis. Contribution of Working Group I to the Fifth Assessment Report of the Intergovernmental Panel on Climate Change, edited by: Stocker, T. F., Qin, D., Plattner, G.-K., Tignor, M., Allen, S. K., Boschung, J., Nauels, A., Xia, Y., Bex, V., and Midgley, P. M., Cambridge University Press, Cambridge, UK and New York, NY, USA, 1535 pp., 2013.

Ismail, O. and Umukoro, G.: Global Impact of Gas Flaring, Energy and Power Engineering, 4, 290–302, https://doi.org/10.4236/ep.2012.44039, 2012.

Julius, O. O.: Environmental impact of gas flaring within Umutebeadi gas plant in Delta State, Nigeria, Archives of Applied Science Research, 3, 272–279, 2011.

Klimont, Z., Kupiainen, K., Heyes, C., Purohit, P., Cofala, J., Rafaj, P., Borken-Kleefeld, J., and Schöpp, W.: Global anthropogenic emissions of particulate matter including black carbon, Atmos. Chem. Phys., 17, 8681–8723, https://doi.org/10.5194/acp-17-8681-2017, 2017.

Li, C., Hsu, N. C., Sayer, A. M., Krotkov, N. A., Fu, J. S., Lamsl, L. N., Lee, J., and Tsay, S.-C.: Satellite observation of pollutant emissions from gas flaring activities near the Arctic, Atmos. Environ., 133, 1–11, https://doi.org/10.1016/j.atmosenv.2016.03.019, 2016.

Liu, Y., Hu, C., Zhan, W., Sun, C., Murch, B., and Ma, L.: Identifying industrial heat sources using time-series of the VIIRS Nightfire product with an object-oriented approach, Remote Sens. Environ., 204, 347–365, https://doi.org/10.1016/j.rse.2017.10.019, 2018.

McEwen, J. D. and Johnson, M. R.: Black carbon particulate matter emission factors for buoyancy-driven associated gas flares, J. Air Waste Manage., 62, 307–321, https://doi.org/10.1080/10473289.2011.650040, 2012.

Nwaichi, E. and Uzazobona, M.: Estimation of the CO2 Level due to Gas Flaring in the Niger Delta, Research Journal of Environmental Sciences, 5, 565–572, https://doi.org/10.3923/jres.2011.565.572, 2011.

Nwoye, C. I., Nwakpa, S. O., Nwosu, I. E., Odo, J. U., Chinwuko, E. C., and Idenyi, N. E.: Multi-Factorial Analysis of Atmospheric Noise Pollution Level Based on Emitted Carbon and Heat Radiation during Gas Flaring, J. Atmos. Pollut., 2, 22–29, 2014.

Obiob, I., Oluwole, A. F., and Akeredolu, F. A.: Non-CO2 gaseous emissions from upstream oil and gas operations in Nigeria, Environ. Monit. Assess., 31, 67–72, 1994.

Ojiijaugo, E., Oduoza, C. F., and Emekwuru, S.: Economics of gas to wire technology applied in gas flare management, Engineering Science and Technology, an International Journal, 19, 2109–2118, https://doi.org/10.1080/10473289.2016.90.012, 2016.

Olivier, J., Janssens Maenhout, G., Muntean, M., and Peters, J.: Trends in global CO2 emissions: 2014 Report, Tech. rep., European Commission, Joint Research Centre (JRC), Institute for Environment and Sustainability (IES), PBL Netherlands Environmental Assessment Agency, the Hague, the Netherlands, 2014.

Onu, P. u., Quan, X., and Ling, X.: Acid rain: an analysis on the cause, impacts and abatement measures Niger Delta perspective, Nigeria, International Journal of Scientific & Engineering Research, 5, 1214–1219, 2014.

Papaillas, G. and Mavroidis, I.: Atmospheric Emissions from Oil and Gas Extraction and Production in Greece, Atmospheric, 9, 152, https://doi.org/10.3390/atmos9040152, 2018.

Qi, L. and Wang, S.: Sources of black carbon in the atmosphere and in snow in the Arctic, Sci. Total Environ., 691, 442–454, https://doi.org/10.1016/j.scitotenv.2019.07.073, 2019.

Quinn, P. K., Shaw, G., Andrews, E., Dutton, E. G., Ruoho-Airola, T., and Gong, S. L.: Arctic haze: current trends and knowl-
edge gaps, Tellus B, 59, 99–114, https://doi.org/10.1111/j.1600-0889.2006.00238.x, 2007.
Rahimpour, M. R. and Jokar, S. M.: Feasibility of flare gas reformation to practical energy in Farashband gas refinery: No gas flaring, J. Hazard. Mater., 209–210, 204–217, https://doi.org/10.1016/j.jhazmat.2012.01.017, 2012.
Schwarz, J. P., Holloway, J. S., Katic, J. M., McKeen, S., Kort, E. A., Smith, M. L., Ryerson, T. B., Sweeney, C., and Peischl, J.: Black Carbon Emissions from the Bakken Oil and Gas Development Region, Environ. Sci. Technol. Lett., 2, 281–285, https://doi.org/10.1021/acs.estlett.5b00225, 2015.
Serreze, M. C. and Barry, R. G.: Processes and impacts of Arctic amplification: A research synthesis, Global Planet. Change, 77, 85–96, https://doi.org/10.1016/j.gloplacha.2011.03.004, 2011.
Siebert, L., Simkin, T., and Kimberly, P.: Volcanoes of the World, 3rd edn., University of California Press, available at: http://www.jstor.org/stable/10.1525/j.ctt1pnqdx (last access: 1 May 2019), 2010.
Stohl, A., Klimont, Z., Eckhardt, S., Kupiainen, K., Shevchenko, V. P., Kopeikin, V. M., and Novigatsky, A. N.: Black carbon in the Arctic: the underestimated role of gas flaring and residential combustion emissions, Atmos. Chem. Phys., 13, 8833–8855, https://doi.org/10.5194/acp-13-8833-2013, 2013.
Uzomah, V. and Sangodoyin, A.: Rainwater chemistry as influenced by atmospheric deposition of pollutants in Southern Nigeria, Environmental Management and Health, 11, 149–156, https://doi.org/10.1108/0956616010321569, 2000.
Venzke, E.: Global Volcanism Program – Volcanoes of the World, v. 4.7.4. Smithsonian Institution, https://doi.org/10.5479/si.GVP.VOTW4-2013, 2013.
Weyant, C. L., Shepson, P. B., Subramanian, R., Cambaliza, M. O. L., Heimburger, A., McCabe, D., Baum, E., Stirm, B. H., and Bond, T. C.: Black Carbon Emissions from Associated Natural Gas Flaring, Environ. Sci. Technol., 50, 2075–2081, https://doi.org/10.1021/acs.est.5b04712, 2016.
Wooster, M. J., Roberts, G., and Perry, G. L. W.: Retrieval of biomass combustion rates and totals from fire radiative power observations: FRP derivation and calibration relationships between biomass consumption and fire radiative energy release, J. Geophys. Res., 110, D24311, https://doi.org/10.1029/2005JD006318, 2005.
Zolfaghari, M., Pirouzfar, V., and Sakhaeinia, H.: Technical characterization and economic evaluation of recovery of flare gas in various gas-processing plants, Energy, 124, 481–491, https://doi.org/10.1016/j.energy.2017.02.084, 2017.