Abstract: Owing to a scarcity of in situ streamflow data in ungauged or poorly gauged basins, remote sensing data is an ideal alternative. It offers a valuable perspective into the dynamic patterns that can be difficult to examine in detail with point measurements. For hydrology, soil moisture is one of the pivotal variables which dominates the partitioning of the water and energy budgets. In this study, nine Irish catchments were used to demonstrate the feasibility of using remotely sensed soil moisture for discharge prediction in ungauged basins. Using the conceptual hydrological model “Soil Moisture Accounting and Routing for Transport” (SMART), behavioural parameter sets (BPS) were selected using two different objective functions: the Nash Sutcliffe Efficiency (NSE) and Coefficient of Determination ($R^2$) for the calibration period. Good NSE scores were obtained from hydrographs produced using the satellite soil moisture BPS. While the mean performance shows the feasibility of using remotely sensed soil moisture, some outliers result in negative NSE scores. This highlights that care needs to be taken with parameterization of hydrological models using remotely sensed soil moisture for ungauged basin.

Keywords: soil moisture; satellite remote sensing; hydrological modelling; ungauged basins

1. Introduction

The conventional ways of monitoring the earth’s hydrological circle depends upon in situ spot measurements. While such measurements absolutely provide information about discrete points within a catchment, they may be inadequate for answering specific spatial-scale research questions. In spite of the saying, “Observation is the foundation of all learning”, observation of nearly all the process components of hydrology is not easy resulting from explicit routine measurements of evaporation, infiltration and subsurface flow and even rainfall, which are difficult to collect in some circumstances. Elevation of water level in the water bodies and through the inference flow in the water bodies is the one quantity that can be measured confidently and which additionally has the merit to be an integrated measurement with respect of the catchment scale. This is probably the reason that human civilizations have relied so greatly on river gauging to recognize, understand and quantify the response of the landscape to rainfall inputs [1]. However, globally, there are an interesting number of catchments with little or no discharge measurements and the number of gauging stations is continuing to decline [2,3]. To help address this lack of information, the International Association of Hydrological Science undertook and completed its science decade on “Predication in Ungauged Basins” (PUB) in 2013 [1]. PUB aimed to answer how to undertake hydrological modelling with consideration of the following: (1) the same location but under
changed circumstances (land use/land cover, observed climate and future climate projection); (2) a different location (gauged/ungauged basins); and (3) different spatial/temporal circumstances (gauged or ungauged catchment under changed conditions).

Ungauged basins usually refer to catchments with a lack of meteorological data (e.g., precipitation and evaporation) and/or hydrological data (i.e., discharge or water level). The lack of either dataset makes calibration of hydrological models impossible or extremely difficult. From here on, ungauged basins/catchments will refer to basins without hydrological data. Remote sensing offers a potential solution to prediction in ungauged basins. Satellite altimetry can measure changes in water levels depending on the size of the water bodies [4,5].

Satellite altimetry is fairly well trusted for use in the calibration of hydrological models as their outputs are similar to traditional in situ measurements; however, currently, only the largest rivers and waterbodies are captured. To capture smaller basins, additional alternatives need to be examined. One method is to examine the fluxes occurring inside the catchment and to try to replicate these. One such flux is the change in soil moisture. Soil moisture is a critical variable of the hydrological cycle, which dominates the partitioning of the mass budget and energy flows between the terrain, ocean and the atmosphere; thus, it is of great importance in the assessment of the different components of the water and energy balance [6]. In situ soil moisture has been employed for improving the estimation of initial conditions in flood modelling [7,8], the calibration of hydrological models [9,10] and the simulation of inundation through data assimilation approaches [11–13].

Soil moisture has been successfully retrieved from both microwave and optical/thermal infrared sensors [14–16]. In fact, global soil moisture products from microwave observations have been developed [17,18], such as the Advanced Microwave Scanning Radiometer-EOS (AMSR-E) [19], the advanced Scatterometer (ASCAT) [20], the Soil Moisture and Ocean Salinity (SMOS) [21], the Soil Moisture Active Passive (SMAP) [22] and the European Space Agency’s Climate Change Initiative (ESA CCI) soil moisture products [23–25]. Several studies have validated different remotely sensed products with in situ measurements to quantify the accuracy of these soil moisture products [25–29]. Remotely sensed soil moisture has also a significant possibility for enhancing river flow predictions in ungauged basins or rarely gauged basins, as demonstrated by [30] with ERS Scatterometer data, and has been shown to improve and support hydrological and climatic predictions [31,32]. Remotely sensed soil moisture products have been used to calibrate hydrological models, from simple water resources accounting models [33] to complex distributed hydrologic models [34,35] and semi-distributed hydrologic models [36,37] in between. Several studies showed that improvement on streamflow simulations were achieved by calibrating hydrologic models against remotely sensed soil moisture data compared with uncalibrated models, though they also showed less improvement than models calibrated against streamflow measurements [33,37].

Three major limitations of satellite soil moisture products were noted by [6]: (1) the sensed soil layer from the satellites is relatively shallow (2–7 cm deep); (2) the currently available products are characterized by the coarse spatial resolution (approximately 25 km); and (3) the quality in regard to certain surface conditions (dense vegetation, frozen soils, snow and mountainous terrain) is relatively low. For these reasons, a number of approaches were developed to deal with them [38] including spatial downscaling techniques [39–41] and simplified methods for root-zone soil moisture estimation from surface measurements [42,43].

In this study, we present the results from using remotely sensed soil moisture for prediction of discharge in Irish catchments. Discharge predictions derived from remotely sensed soil moisture will be compared to those derived from calibration using in situ discharge. The main objectives of this paper are to (1) investigate the potential of using remotely sensed soil moisture for ungauged basins in Ireland and (2) to investigate the change in accuracy between remotely sensed and in situ derived predictions.
2. Data and Methodology

2.1. SMART Model

A novel model—the soil moisture accounting and routing for transport model—was developed and informed with regard to existing model structures, especially the pathway separation of NAM (NedborAfstromnings Model) [44] and the soil layers of the SMARG (model [45,46]).

The SMART model is a lumped model forced with precipitation and potential evapotranspiration, and the soil components are represented by six layers of equal depth with the total depth/capacity of the six layers equal to $Z$. If the soil moisture capacity of a layer is exceeded, excess moves to a deeper layer if feasible or is intercepted by one of the flow pathways other than overland flow [46]. Five different flow pathways are simulated in the model: (1) overland flow (depending on the direct runoff parameter ($H$) and the relative soil moisture content), (2) drain flow, (3) interflow (depending on soil moisture excess and outflow from the soil layer), (4) shallow groundwater and (5) deep groundwater. These five pathways are routed initially through four routing components and finally a river routing parameter (Figure 1). All parameters and recommended ranges are found in Table 1.

![Figure 1. Schematic representation of the Soil Moisture Accounting and Routing for Transport (SMART) model [45].](image)

| Parameter | Description                              | Range       |
|-----------|------------------------------------------|-------------|
| T         | Areal Rainfall Correction coefficient ($T$) | 0.9–1.1     |
| C         | Evaporation decay parameter ($C$)         | 0–1         |
| H         | Quick runoff coefficient ($H$)            | 0–0.3       |
| S         | Drain flow parameter ($S$)                | 0–1         |
| D         | Soil outflow coefficient ($D$)            | 0–0.013     |
| Z         | Effective soil depth (mm)                | 15–150      |
| SK        | Surface routing parameter (hours)         | 1–240       |
| FK        | Interflow routing parameter (hours)       | 48–1440     |
| GK        | Groundwater routing parameter (hours)     | 1200–4800   |
| RK        | River routing parameter (hours)           | 1–96        |

Table 1. SMART model parameters, descriptions, units and recommended ranges.
2.2. Study Catchments

Nine Catchments in Ireland were selected with the aim of representing the various geography, topography and meteorology conditions. Name, gauged area and location of these catchments are presented in Figure 2 and Table 2. The gauging locations chosen are the furthest downstream gauges, where discharge measurements are available for the period of this study. These study catchments were also used in previous studies looking at hydrological modelling in Ireland either in full [45] or a subset [46,47].

Table 2. Study catchments including catchment size, nearest synoptic station and discharge gauge.

| Catchment Name | Gauged Area (km²) | Synoptic Station Name | Discharge Gauge Number | Discharge Gauge Name       |
|----------------|-------------------|-----------------------|------------------------|---------------------------|
| Barrow         | 2433              | Casement              | 14,018                 | Royal Oak                 |
| Blackwater     | 2315              | Cork Airport          | 18,002                 | Ballyduff                 |
| Boyne          | 2467              | Dublin Airport        | 7012                   | Slane Castle              |
| Dee            | 305               | Dublin Airport        | 6013                   | Charleville               |
| Feale          | 648               | Valentia Observatory  | 23,002                 | Listowel                  |
| Moy            | 1908              | Belmullet             | 34,001                 | Rahans (Goswami and O’Connor) |
| Nore           | 2380              | Casement              | 15,006                 | Brownsbarn                |
| Suck           | 1215              | Shannon Airport       | 26,007                 | Bellagill                 |
| Suir           | 1565              | Shannon Airport       | 16,009                 | Cahir Park                |

Figure 2. Study catchments with elevation.
2.3. Meteorological Data

To run the SMART model, both rainfall and potential evapotranspiration (PET) are required. Daily rainfall data for relevant rain gauges were obtained from Met Éireann, the Irish meteorological office. These daily point rainfall estimates were converted into catchment average rainfall using the Thiessen Polygon method. For the nine study catchments, the annual average rainfall varied from 864.2 mm to 1392.7 mm.

Potential evapotranspiration was also obtained from Met Éireann. Met Éireann estimates PET using the FAO Penman–Monteith formula. However, unlike for rainfall, there are fewer locations where PET is measured. Therefore, PET for each catchment was estimated by taking the values from the nearest synoptic station. Table 2 lists the chosen synoptic stations for each catchment. Annual average PET across the six stations varies from 526.6 mm to 581.1 mm.

2.4. Discharge

Daily mean discharge records from 1990 to 2005 were collected from the furthest downstream active gauge of each catchment. This data was obtained from the Irish Office of Public Works (OPW), and this data is freely available at https://waterlevel.ie/hydro-data/. There was on average 5.3% missing data from all nine gauges with a gauge number 26,007 on Suck recording the largest amount of missing data with 22.24%.

2.5. Soil Moisture

The soil moisture data were collected from the Global ASCAT Surface Soil Moisture Data Service belonging to the GEO Department of TU Wien, who have been investigating and constantly improving algorithms for retrieving soil moisture data from C-band scatterometer measurements since the mid-1990s [48].

The data of surface soil moisture are retrieved from the radar backscattering coefficients measured by the ASCAT (advanced scatterometer) instrument on board the MetOp-A satellite using a change detection method, developed at the Research Group Remote Sensing, Department of Geodesy and Geoinformation (GEO), Vienna University of Technology (TU Wien).

The soil moisture product represents the water content in the upper soil layer (<2 cm) and in relative units between totally dry conditions (0%) and total water capacity (100%). The time series are available on a discrete global grid (DGG) with a spatial resolution of 25 km (grid spacing 12.5 km). The temporal sampling rate is irregular (every 1–2 days) and depends on the latitude. More detail on the retrieval methodology is described in [49].

2.6. Experimental Design and Hypothesis Testing

After collecting the data, 10,000 parameter sets were generated through the Latin hypercube sampling method with the assumption of uniform distribution of the parameter space. The SMART model was then run at a daily timestep using these parameter sets to generate 10,000 unique simulations for each of the nine catchments included in this study. For each simulation, the discharge (m$^3$/s) and the state of each of the six soil moisture layers (m$^3$) were simulated. To investigate the potential benefit of remotely sensed soil moisture in ungauged catchments, for each catchment, the simulations were compared separately to both observed discharge and remotely sensed soil moisture to determine behavioural parameter sets (BPSs). In this study, BPSs were determined as the top 5% best performing parameters sets in terms of comparison between the simulated and observed discharges and the Nash Sutcliffe efficiency [50] and in terms of $R^2$ comparison between the simulated state of the top soil layer and the remotely sensed soil moisture. A very similar methodology has previously been used [46] and closely resembles the Generalized Likelihood Uncertainty Estimation (GLUE) methodology [51].
The simulated period from 1990 to 2005 was separated into three periods: model warm-up (1 January 1990 to 31 December 1990), calibration (1 January 1991 to 30 June 1998) and validation (1 July 1998 to 31 December 2005). The BPSs were chosen using the calibration period and one of the objective functions and compared to both objective functions in validation. This was done for both objective functions separately.

Using the experiment design above, we aim to test three hypotheses during this study:

1. Calibration using only remotely sensed soil moisture would perform worse than calibrating to discharge data.
2. Remote sensed soil moisture will only be able to provide small benefits to simulate discharge in ungauged basins, especially in the case of Ireland.
3. A clear relationship between catchment size and model performance will be exhibited when using the \( R^2 \) BPS.

3. Results

In total, four different sets of results were produced:

- Calibration using Nash Sutcliffe Efficiency (NSE) and validated using NSE
- Calibration using NSE and validated using \( R^2 \)
- Calibration using \( R^2 \) and validated using \( R^2 \)
- Calibration using \( R^2 \) and validated using NSE

3.1. Calibration by Nash Sutcliffe Efficiency (NSE)

Figure 3 shows the summary of the behavioural parameter set (BPS) based on NSE for the calibration and validation period with Table 3 giving the mean, standard deviation and maximum NSE and \( R^2 \) for each catchment in calibration and validation. The results show that the SMART model can reproduce the observed discharge to obtain high NSE values in most catchments and that there is a very small standard deviation across the BPS. The one exception is the Feale catchment for which the median NSE is lower than the others in both calibration and validation. However, there is a small difference if only the best simulation is chosen.

|                | NSE Calibration (Validation) | R² Calibration (Validation) |
|----------------|-----------------------------|-----------------------------|
|                | Boyne Barrow Nore Boyne     | Boyne Barrow Nore           |
| Mean           | 0.792 (0.824) 0.83 (0.842)  | 0.627 (0.472) 0.591 (0.452) |
| Standard Deviation | 0.025 (0.028) 0.023 (0.03) | 0.029 (0.042) 0.038 (0.045) |
| Max            | 0.882 (0.911) 0.92 (0.929)  | 0.674 (0.578) 0.656 (0.552) |
|                | Blackwater Moy Suir Blackwater | Moy Suir               |
| Mean           | 0.838 (0.811) 0.896 (0.91)  | 0.595 (0.404) 0.607 (0.434) |
| Standard Deviation | 0.024 (0.041) 0.013 (0.019) | 0.034 (0.039) 0.035 (0.03)  |
| Max            | 0.906 (0.914) 0.939 (0.952) | 0.663 (0.49) 0.66 (0.51)   |
|                | Suck Feale Dee Suck         | Feale Dee                |
| Mean           | 0.903 (0.915) 0.614 (0.593) | 0.629 (0.447) 0.569 (0.365) |
| Standard Deviation | 0.016 (0.018) 0.064 (0.072) | 0.045 (0.042) 0.024 (0.024) |
| Max            | 0.943 (0.95) 0.809 (0.817)  | 0.698 (0.546) 0.667 (0.438) |

The results for \( R^2 \) in both the calibration and validation periods using the NSE BPS show similar patterns as the NSE results. The average \( R^2 \) value across all catchments was 0.604 in the calibration period.
and 0.439 in the validation period. As with the NSE results, Suck performed best for the calibration period with a $R^2$ value of 0.629 and the Feale catchment performed the worst in both periods. As with the NSE results, the results of the $R^2$ analysis, using the NSE BPS, show that catchment size does not have any role in the performance. This is seen in Figure 3 where the catchments are shown in order of catchment size with the Boyne being the largest and the Dee being the smallest.

**Figure 3.** Box plots of NSE and $R^2$ performance using the NSE BPS, (a) NSE results for Calibration; (b) NSE results for Validation; (c) $R^2$ results for Calibration; and (d) $R^2$ results in Validation.

Figure 4 shows the mean and standard deviation of all NSE BPS simulations for all nine study catchments compared to in situ observations. With the exception of Feale, it is clearly seen that the mean of the NSE BPS simulations is able to closely reproduce the temporal pattern observed in the in situ observations. This is the case for both high and low flow periods. This is not unexpected as the use of NSE as an objective function is widely used, though it does place more emphasis on high flows.

### 3.2. Calibration by $R^2$

The second BPS was obtained by a $R^2$ comparison between the top soil moisture layer in the SMART model and the remotely sensed soil moisture product. Figure 5 and Table 4 show the performance of the $R^2$ BPS in relation to NSE and $R^2$ in both the calibration and validation periods. The results are very similar to those using the NSE BPS with an average $R^2$ of 0.653 in the calibration period and 0.488 in the validation period. As with the previous results, Suck is the best performing catchment in the calibration period; however, Feale is no longer the worst catchment in both periods and is the fifth best catchment in the calibration period but the worst in the validation period. As expected, comparing the $R^2$ results for both BPS, the spread in performance as measured by the standard deviation is far smaller with the $R^2$ BPS with an order of magnitude difference in the calibration period.

The results for NSE in both the calibration and validation periods using the $R^2$ BPS show a very similar pattern to those using the NSE BPS. The majority of catchments behave very similarly and produce very comparable NSE scores, with Feale being an exception again. The average of the mean performance across all catchments is smaller than those using the NSE BPS at 0.630 and 0.624 in the calibration and validation periods, respectively. However, for the majority of the catchments, with the exception of Feale, there are a number of outliers for which the NSE scores are less than 0. This indicates that, for some simulations, the performance is worse than the mean discharge. If we ignore outliers, the NSE performance is always greater than 0. Looking at the interquartile ranges (green area of box plots), the results show that 50% of the $R^2$ BPS produce NSE scores greater than 0.50 in both the calibration and validation periods.
Figure 4 shows the mean and standard deviation of all NSE BPS simulations for all nine study catchments compared to in situ observations. With the exception of Feale, it is clearly seen that the mean of the NSE BPS simulations is able to closely reproduce the temporal pattern observed in the in situ observations. This is the case for both high and low flow periods. This is not unexpected as the use of NSE as an objective function is widely used, though it does place more emphasis on high flows.

**Figure 4.** Simulated hydrographs using the NSE BPS compared to in situ observations for all catchments: (a) Boyne; (b) Barrow; (c) Nore; (d) Blackwater; (e) Moy; (f) Suir; (g) Suck; (h) Feale; and (i) Dee. The mean of BPS corresponds to the mean of all BPS simulations.
Table 4. R² summary statistics (NSE and R²) for calibration and validation using the R² BPS in the format calibration (validation).

| NSE  | Calibration | Validation |
|------|-------------|------------|
|      | Boyne      | Barrow     | Nore       | Boyne      | Barrow     | Nore       |
| Mean | 0.650 (0.661) | 0.679 (0.676) | 0.661 (0.634) | 0.661 (0.521) | 0.639 (0.504) | 0.634 (0.513) |
| Standard Deviation | 0.138 (0.150) | 0.172 (0.177) | 0.132 (0.146) | 0.005 (0.017) | 0.006 (0.016) | 0.007 (0.014) |
| Max  | 0.882 (0.885) | 0.92 (0.928) | 0.906 (0.885) | 0.679 (0.588) | 0.662 (0.561) | 0.661 (0.564) |

| R²   | Calibration | Validation |
|------|-------------|------------|
|      | Blackwater  | Moy        | Suir       | Blackwater  | Moy        | Suir       |
| Mean | 0.683 (0.655) | 0.691 (0.73) | 0.743 (0.732) | 0.64 (0.459) | 0.656 (0.485) | 0.66 (0.487) |
| Standard Deviation | 0.115 (0.12) | 0.135 (0.135) | 0.133 (0.128) | 0.007 (0.01) | 0.006 (0.015) | 0.004 (0.007) |
| Max  | 0.901 (0.9) | 0.922 (0.943) | 0.94 (0.912) | 0.672 (0.49) | 0.682 (0.548) | 0.68 (0.513) |

|      | Suck       | Feale      | Dee        | Suck       | Feale      | Dee        |
| Mean | 0.651 (0.637) | 0.297 (0.268) | 0.614 (0.627) | 0.692 (0.515) | 0.651 (0.402) | 0.645 (0.507) |
| Standard Deviation | 0.161 (0.166) | 0.119 (0.119) | 0.15 (0.163) | 0.005 (0.012) | 0.006 (0.016) | 0.007 (0.015) |
| Max  | 0.925 (0.939) | 0.717 (0.717) | 0.88 (0.915) | 0.712 (0.551) | 0.678 (0.457) | 0.674 (0.561) |

Figure 5. Boxplot of NSE and R² performance using the R² BPS, (a) NSE results for Calibration; (b) NSE results for Validation; (c) R² results for Calibration; and (d) R² results in Validation.

Figure 6 shows the mean and standard deviation of all R² BPS simulations for all nine study catchments compared to in situ observations. A similar pattern to the hydrographs shown in Figure 4 can be seen in the mean and standard deviation of the R² BPS simulations, with the Feale catchment performing worse. For all other catchments, it is clearly seen that the simulations perform better for low flow periods than for higher flow periods. This is clearly visible for the larger catchments, such as Nore or Blackwater, where the mean and standard deviation of the R² BPS are unable to match the observed peak flows but are able to capture the low flow periods.
Figure 6. Simulated hydrographs using the R² BPS compared to in situ observations for all catchments: (a) Boyne; (b) Barrow; (c) Nore; (d) Blackwater; (e) Moy; (f) Suir; (g) Suck; (h) Feale; and (i) Dee. The mean of BPS corresponds to the mean of all BPS simulations.
4. Discussion

4.1. Hydrograph Comparison

It is clear from Figures 4 and 6 that the NSE BPS can match the observed temporal variation in discharge better than the R² BPS. Both BPSs were able to capture low flow periods within the one standard deviation of the mean of all simulations, while the R² BPSs were unable to match either the NSE BPS or the observations during high flows. However, this is expected as the NSE BPSs were determined by comparing simulated discharge to in situ observations while the R² BPSs were determined by comparing the top soil layer in the SMART model to remote sensed soil moisture estimates. Therefore, in the case of the R² BPS, no information the hydrograph response was used.

4.2. Hypothesis Tests

Prior to analysing the results, it was hypothesized that calibration of the SMART model using only soil moisture from satellite remote sensing would (1) perform worse than calibrating to discharge data; (2) provide only a small benefit to simulate discharge in ungauged basins, especially in the case of Ireland; and (3) exhibit a clear relationship with catchment size.

By comparing the results of the two sets of behavioural parameters, it is clear that hypothesis 1 is true but not to the scale expected. A far larger difference in performance was expected, as the catchments studied were relatively small compared to the spatial footprint of the soil moisture data. The spatial footprint of the soil moisture product is 625 km², while the catchments studied varied between 305 km² and 2467 km². Using the NSE BPS, the average of NSE performance across all catchments was 0.815 and 0.820 in the calibration and validation periods, respectively, while the performances across the same two period using the R² BPS were 0.630 and 0.624. While this clearly shows that calibrating with discharge data is superior than calibrating with remotely sensed soil moisture data, it does show that, for the nine catchments in this study, soil moisture is a very important model state in simulating discharge. This is also supported by Figures 4 and 6.

This also disproves our second hypothesis that remotely sensed soil moisture would only have a small benefit in ungauged catchments. Across all the catchments calibrated using the R² BPS, the SMART model was able to produce simulated hydrographs that produced good NSE scores. The calibration of the SMART model with only remotely sensed soil moisture was able to simulate realistic hydrographs, as shown in Figure 6, though higher flows may be underestimated if only remotely sensed soil moisture for calibration is used and though in an ungauged catchment this might be the only information available. There were a small number of behavioural parameter sets that resulted in negative NSE values, indicating that, for these small number of simulations, the performances were worse than the mean discharge; however, these did not result in the mean and standard deviations of all R² BPS simulation being unusable.

As previously stated, it was expected that there would be a clear relationship between catchment size and the benefit of remotely sensed soil moisture. It was expected, based off a previous study [52], that the larger catchments would produce better results than smaller catchments in the study. The results would indicate that there is no relationship between remote sensed soil moisture, catchment size and model performance. From the results, it is clear that the result using the R² BPS follow the same pattern as the NSE BPS. Neither set exhibits any relationship with catchment size, and both sets performed poorly for the same catchment (Feale).

4.3. Limitation and Future Work

In this study, nine catchments were chosen for the purpose of this research. These were chosen to be representative of the geography, meteorology and topography of Ireland. No strong inference to the value of remotely sensed soil moisture for smaller catchments can be made from this study, as the smallest catchment chosen was Dee for which catchment area is 305 km². However, a previous study [52] noted that it is better to choose catchments where the area is greater than one pixel of the remote sensing product. Four catchments (Boyne, Barrow, Nore and Blackwater) were chosen in which
the catchment areas are greater than 2000 km² and these cover the majority of the large Irish rivers. Catchments that are heavily influenced by Karst features were also excluded from this study. Any further study should try to address the limitations mentioned and investigate if the value of remotely sensed soil moisture is consistent throughout larger catchments, if including more layers in the analysis would improve the results, if including existing methods of estimating soil moisture to depths of 1 m have any benefit and if the results are model dependent.

5. Conclusions

Nine gauged catchments across Ireland with catchment areas ranging from 305 km² to 2467 km² were chosen to investigate the feasibility of discharge prediction in ungauged basins using only remotely sensed soil moisture. A lumped conceptual hydrological model, the Soil Moisture Accounting and Routing for Transport (SMART) was calibrated and validated by observed discharge and remotely sensed soil moisture using the Nash Sutcliffe Efficiency (NSE) and coefficient of determination (R²) respectively to select Behavioural Parameter Sets (BPS) for each objective function. Using these BPS, three hypotheses were tested: (1) calibration using only remotely sensed soil moisture would perform worse than calibrating to discharge data; (2) remote sensed soil moisture will only be able to provide small benefits to simulate discharge in ungauged basins, especially in the case of Ireland; and (3) a clear relationship between catchment size and model performance will be exhibited when using the R² BPS.

From the results, hypothesis 1 was proven true but to a smaller degree than expected. Calibrating the SMART model using remotely sensed soil moisture was able to simulate hydrographs that captured the low flows though underestimated the peak flow while still obtaining good NSE values (on average 0.630 across all nine catchments in the calibration period). Hypothesis 2 was disproven. Simulated hydrographs using parameter only estimated from remotely sensed soil moisture was able to closely reproduce the temporal variation of the in situ observations, though performing better at low flow periods compared to peak flows. Finally, the third hypothesis was disproven. It was assumed that a relationship in performance and catchment size would be visible in the results, especially when using remotely sensed soil moisture. However, this relationship was not apparent with either objective function or in the simulated hydrographs.

Care must be taken using only remotely sensed soil moisture for calibration as it provides no information on the hydrological response of the river, which is necessary to accurately capture peak flow. Nonetheless, this study found that remotely sensed soil moisture is feasible for the parameterization of hydrological model for use in ungauged basins.

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