Maturity, growth and natural mortality rate of the introduced fish *Parachromis managuensis* (Perciformes: Cichlidae) in the semiarid region of Brazil

Maturidade, crescimento e taxa de mortalidade natural do peixe introduzido *Parachromis managuensis* (Perciformes: Cichlidae) na região semiárida do Brasil

Amanda Graziele Araújo Resende¹,²* Elton José de França³ and Francisco Marcante Santana¹,²

¹Laboratório de Dinâmica de Populações Aquáticas – DAQUA, Unidade Acadêmica de Serra Talhada – UAST), Universidade Federal Rural de Pernambuco – UFRPE, Avenida Gregório Ferraz Nogueira, s/n, José Tomé de Souza Ramos, CEP: 56909-535, Serra Talhada, PE, Brasil
²Programa de Pós-graduação em Biodiversidade e Conservação – PGBC, Unidade Acadêmica de Serra Talhada – UAST, Universidade Federal Rural de Pernambuco – UFRPE, Avenida Gregório Ferraz Nogueira, s/n, José Tomé de Souza Ramos, CEP: 56909-535, Serra Talhada, PE, Brasil
³Laboratório de Ecologia e Sistemática de Peixes – LAESP, Unidade Acadêmica de Serra Talhada – UAST, Universidade Federal Rural de Pernambuco – UFRPE, Avenida Gregório Ferraz Nogueira, s/n, José Tomé de Souza Ramos, CEP: 56909-535, Serra Talhada, PE, Brasil
*e-mail: amandagraziele019@gmail.com

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Abstract: Aim: The present study aimed to estimate the sexual maturity, growth and mortality rates of the jaguar guapote, *Parachromis managuensis*, an introduced fish species in the Brazilian semiarid region. Methods: Specimens were collected between December 2015 and November 2016, in a dam at Serra Talhada municipality, Pernambuco, where the species was abundant. Results: In total, 520 individuals were analyzed, with total lengths (TL) varying between 23 and 240 mm. Length at sexual maturity varied between 130.87 and 140.68 mm _TL_ (Mean = 135.78 mm _TL_). Through the modal progression of monthly frequency distributions, the following growth parameters were obtained according to the von Bertalanffy model: _L_∞ = 341.40 mm, _k_ = 0.840 and _t_0 = -0.035 year⁻¹. Ages varied from 0.05 to 1.41 year, and an age at first maturity between 0.48 and 0.53 year (Mean = 0.51 year), with the largest part of the sample (72.64%) made of adult individuals. Natural mortality rates (M) were estimated from three different methods, varying between 1.53 and 3.04, with a mean of 2.27. Conclusions: *P. managuensis* shows fast growth and maturation and high natural mortalities, which grants it biological advantages when compared to native species from the same trophic level in the area.

Keywords: Jaguar Guapote; Northeastern Brazil; length-weight relationship; age structure.

Resumo: Objetivo: O presente estudo teve como objetivo estimar a maturidade sexual, o crescimento e as taxas de mortalidade do peixe introduzido na região do semiárido brasileiro, o jaguar guapote, *Parachromis managuensis*. Métodos: Os exemplares foram coletados entre os meses de Dezembro de 2015 e Novembro de 2016, em um açude do município de Serra Talhada, Pernambuco, onde a espécie é abundante. Resultados: Um total de 520 indivíduos foram analisados,
Among the anthropic actions, the introduction of non-native species has led to ecological significant events, biological invasions, which can affect biodiversity from the genetic to the landscape level (Vitule, 2009; Vitule & Prodocimo, 2012). These species introductions and translocations are among the main adversities that affect freshwater fish conservation, since non-native species compete for space and food. Therefore, they cause imbalances in ecological niches, altering the food chain dynamics and, consequently, nutrient transfer (Collares-Pereira & Cowx, 2004; Amador et al., 2009).

In Brazil, several non-native fish species have been reported in different environments (Bittencourt et al., 2014; Gubiani et al., 2018; Patoka et al., 2018). Among these species is the jaguar guapote, *Parachromis managuensis* (Gunther, 1867), a member of the Cichlidae Family. Native to the rivers of Central America, between Honduras and Costa Rica (Bussing, 2002), it has been introduced in many countries, such as Brazil (Agasen et al., 2006; Rosana et al., 2006; Barros et al., 2012; Hamiyati et al., 2019), including the Northeastern semiarid, where species introduction may have been accidental (França et al., 2017; Ramos et al., 2018). *P. managuensis* is widely distributed in Brazil, with reports of its use in pisciculture, fish stocking and aquarium activities (Barbosa et al., 2006; Magalhães & Jacobi, 2010; Barros et al., 2012). This species is described as potentially invasive, displaying a series of key characteristics such as fast maturity, resistance to environmental changes, besides displaying a territorial aggressive behavior, with a mainly piscivorous carnivorous diet (Agasen et al., 2006; Barbosa et al., 2006), which can make it successful in competition with native fishes. For these reasons, it is listed on the invasive species database of the Inter-American Biodiversity Information Network (2019). For these reasons, given adequate environmental conditions, *P. managuensis* can establish itself and cause ecological damages to native species.

The caatinga is a biome exclusive of Brazil and part of this biological heritage is found only in Northeast of country. It is a heterogeneous biome, with great diversity of fauna and flora, and with a high degree of endemism of species. Despite its biological importance, the caatinga still has a low priority and insufficiency in terms of scientific studies (Leal et al., 2005). The Caatinga although exhibiting a dry climate, harbors many springs, such as lakes, dams and rivers, where a large diversity of species dwells. Studies on this diversity revealed that the semiarid region has 386 fish species, out of which 371 are native, 203 are considered endemic to the ecoregions where the Caatinga occurs, 15 are introduced from other basins, and 33 are threatened at some level (Lima et al., 2017). This fact demonstrates the urgency of studies that clarify the biological consequences of non-native species to native and endemic species in this region, which is already in need of studies and has been extensively devastated.

In this context, studies that contribute to the knowledge on population dynamics of non-native fishes aid in the development of risk analyses for these species. Thus, studies on age and growth are important to subsidize data on species population dynamics, since information regarding age structure of a fish population favours the interpretation of growth and survival variability, enabling the comprehension of factors that condition recruitment success (Jones, 1992). For these reasons, this study aimed to determine the population parameters of sexual maturity, growth and natural mortality rates of *P. managuensis* in a reservoir of the Brazilian semiarid region.

## 2. Material and Methods

The study site is at the Brazilian semiarid region, with shows particularities, such as the high spatial and temporal variation in precipitation, high
evapotranspiration potential, a deficit in water balance, and deciduous vegetable cover known as “Caatinga” (Barbosa et al., 2012; Moura et al., 2007). Sampling took place at the Saco I dam, with a water capacity of 36 million m$^3$, located at the Serra Talhada municipality, in the sertão mesoregion of Pernambuco state. The study site belongs to the Pajeú river basin, considered the state’s largest, with an area of 16685.53 km$^2$ and an intermittent river regime (APAC, 2018).

Specimens of *P. managuensis* (Figure 1) were collected monthly at the Saco I dam, from December 2015 to November 2016, with the use of 1.5 and 2.2m cast nets with mesh openings from 15 to 20 mm, and 5 – 30m trawl nets with a 5mm mesh. For each captured specimen, total (TL) and standard (SL) lengths were measured, as well as total weight (TW), with the use of a scale with a 0.0001g precision.

The weight-length relationship (*TW*-TL) for *P. managuensis* was estimated following the potential equation $TW = a^*TL^b$, where $a$ is the coefficient of proportionality and $b$ is the angular coefficient related to the species way of growth (Le Cren, 1951). The weight-length relationship analyzes the differences in growth related to the increase in weight or length, and can assume isometric or allometric characteristics. The existence of isometry ($b = 3$), occurs when the increase in weight accompanies the growth in length, if not, the relationship is allometric, can be positive when the growth in weight is greater than length ($b > 3$) or negative when the growth in length is greater than weight ($b < 3$) (Froese, 2006). The angular coefficient obtained was submitted to a Chi-square test ($\chi^2$), aiming to observe differences of the $b$ value calculated with 3 (isometry). The linear relationship between total and standard lengths ($TL = a^*b^*SL$) was also calculated for the entire sample.

The length at maturity was estimated according to two methods. The first method consists of gonad macroscopic analyses, that were performed to identify specimens sex and maturity (young or adults), according to maturity scale for fishes described by Vazzoler (1996). These information allowed to estimate the length at which 50% of individuals are adults ($L_{50}$) according to the equation by King (1995): $P = 1/[1 + e^{a^*(TL-L_{50})}]$, where $P$ is the proportion of adult individuals per total length (TL) class and $a_0$ is the equation parameter.

The second method used to determine length at sexual maturity was a polyphasic model divided in two stanzas, in which each stanza describes by a power function. The length-weight relationship and a logistic function for the $a$ and $b$ parameters of Huxley (1924) was used to describe relative growth and a change in growth pattern (stanza) in the life history of species. The power function: $y = f(a)x^{f(b)}$, was divided in two stanzas (1 and 2 for each length – weight coefficients $a$ and $b$): $f(a) = a_1 + (a_2 - a_1)/(1 + e^{Rsc(a)x(Lsc-L_0)})$ and $f(b) = b_1 + (b_2 - b_1)/(1 + e^{Rsc(b)x(Lsc-L_0)})$, where, $R_{sc}(a/b)$ is the rate of coefficient changes between stanza 1 and 2, and $L_0$ is the length at the stanza changing point, which was considered as the length at maturity (Fontoura et al., 2010; Freitas et al., 2016). To estimate the parameters for the two methods, functions were adjusted and parameters estimated through the Solver routine available on the Microsoft Excel software using minima square, and the bootstrap analysis to calculate confidence intervals.

Growth parameters of the von Bertalanffy (1938) model were estimated from the following equation: $L_t = L_{\infty} \left(1 - e^{-kt}\right)$, where $L_t$ is the length at age $t$; $L_{\infty}$ is the species asymptotic or theoretical maximum length; $k$ is the growth coefficient and $t_0$ is the theoretical age at length zero. To estimate these parameters, the monthly modes from the length frequency distribution were used through software ELEFAN I (Eletronic Lengths-Frequency Analysis) (Pauly et al., 1981), through a modal progression analysis, performed on the FISAT software (FAO ICLARM Stock Assessment Tools) (Gayanilo et al., 1997). The value for $t_0$ was estimated based on the equation described by Pauly (1979):

$$
\log(-t_0) = -0.3922 - 0.2752 x \log(L_{\infty}) - 1.038 x (k) .
$$

The population age structure was estimated based on its total length (TL), from the inverted Von Bertalanffy equation:

$$
t(L) = t_0 - \frac{1}{k} \ln \left(1 - \frac{TL}{L_{\infty}}\right).
$$

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Figure 1. Specie *P. managuensis* collected in at the Saco I dam, Pernambuco-Brazil.

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Natural mortality rates ($M$) for the species were estimated based on three methods:
Hoennig (2005): $\ln M = 1.46 - 1.01 \times \ln t_{\text{max}}$
Pauly (1980):

$$\ln \ln M = -0.0066 - (0.279 \ln L_\infty) + (0.6543 \ln k) + (0.4634 \ln T)$$

Rikhter & Efanov (1976): $M = \frac{L_{521}}{t_{50}^{0.72}} - 0.155$

In which, $t_{\text{max}}$ is the highest age in the sample; $T$ is the average water temperature (°C), and $t_{50}$ corresponds to the average age at first maturity.

3. Results

In total, 520 specimens of $P. managuensis$ were sampled, with the highest total length frequencies between classes of 130 and 150 mm. From the specimens sampled, 141 were female (68 - 240 mm $TL$, $x = 153.71$ mm $TL$, SD = 21.73 mm $TL$), 95 were male (78 - 230 mm $TL$, $x = 145.87$ mm $TL$, SD = 28.34 mm $TL$) and 284 could not have their sex identified, especially those smaller than 100 mm $TL$, which were considered immature (Figure 2). Due to the low number of identified adult by sex and the difficulties in determining sex for small individuals, the analysis was realized with grouped data, with both sexes together.

Total weight of the sample varied between 0.34 and 160.73 g. The weight-length relationship yielded the following equation: $TW = 0.00003 TL^{2.9004} \ (r^2 = 0.9868)$ (Figure 3A). When testing for possible differences between allometry coefficients by $\chi^2$, a significant difference was noticed ($P = 0.0033$), confirming the negative allometry for the sample. The total length-standard length relationship was described by the equation: $TL = 1.9181 + 1.2657 \times SL \ (r^2 = 0.9888)$ (Figure 3B).

The $L_{50}$ estimated through the model by King (1995) was 140.68 mm $TL$ (95% confidence interval = 137.58/144.45; SD = 1.81 mm) (Figure 4A), while the $L_{50}$ resulted in 130.88 mm $TL$ (95% confidence interval = 129.78/134.56; SD = 1.21 mm) (Figure 4B) with the method described by Freitas et al. (2016). Length at maturity for the species was considered as the average length between both methods (135.78 mm $TL$).

Through the modal progression analysis of monthly length frequency distributions, two six-monthly cohorts were identified (Figure 5). Maximum theoretical length estimated for the species was 341.4 mm $TL$, with a growth coefficient of 0.84. The estimated $t_{0}$ was -0.035 year. With these parameters, the species reaches 123.65 mm during the first 6 months of life, with a monthly growth rate of 20.61 mm, decreasing to 12.45 mm/month afterwards between 6 months and 1 year and to 8.18 mm/month between the ages of 1 and 1.5 years (Figure 6).

Age structure of $P. managuensis$ varied between 0.05 and 1.41 year ($x = 0.53$, SD = 0.23 year), with the majority of individuals aged between 0.4 and 0.8 year (Figure 6). Age at sexual maturity varied between 0.48 and 0.53 year ($x = 0.51$ year, SD = 0.03 year), and fishing recruitment happens at 0.7 year, with the majority of the sample (72.64%) being comprised of adult individuals (Figure 7).

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Figure 2. Total length frequency distribution for female (black), male (gray) and unidentified (white) specimens of $P. managuensis$ at the Saco I dam, Pernambuco-Brazil

Figure 3. Total weight-total length relationship (A) and total length-standard length relationship (B) for $P. managuensis$ at the Saco I dam, Pernambuco-Brazil
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Natural mortality rates ($M$) varied from 1.53 to 3.04, with a mean of 2.27 (Table 1), compatible with fast-growing, early-maturing, short-lived fish species.

4. Discussion

The population of $P. managuensis$ in the Saco I dam demonstrated adaptation to the biological conditions of the study site, presenting growth and reproduction strategies consistent with invasive species, establishing themselves in the place, since it does not present a natural predator, and would not have to theoretically compete for space and food. Conditions that favor its development and propagation, with behavior similar to that already registered in the literature in other countries that was introduced. The rapid growth strategy, associated with the annual average of the local temperature, made it possible to sample individuals with shorter lengths, but physiologically active for reproduction. As noted, Saco I dam’s $P. managuensis$ population showed lower lengths (169 mm SL) than those reported by Bussing (2002) in its native environment of 220 mm standard length. The same happens when comparing it to other non-native cichlid species (Table 2). Consequently, existing variations in length distributions can be influenced by the time since introduction and the characteristics that make up the ecosystems where species were analyzed (Gomiero & Braga, 2003).
The kind of allometry verified for the species through the weight-length relationship contributes to the knowledge on its morphology, providing parameters for comparison with other populations (Montenegro et al., 2011). The type of allometry verified indicates a greater increase in length than in weight, providing evidence of growth strategy, which may be related to the stages of maturity. There is no set pattern, in which the growth constant most likely varies depending on environmental conditions, available resources, and the species biology.

Knowledge regarding average length at first maturity ($L_{50}$) for the species is among the information that contribute to understanding the reproductive strategy, as well as management strategies and minimum capture lengths (Gurgel et al., 2011). Based on that, the estimated length at maturity for *P. managuensis* is similar to that described by Farid et al. (2000) for *Parachromis dovii* (138 mm TL) when using the same method in Costa Rica, and the estimated for *P. managuensis* in Indonesia (136.4 mm TL), obtained based on the Spearman-Karber method in Udupa (1986), which considers maturing individuals as mature (Hamiyati et al., 2019). The size of the first maturation varies between regions, motivated by environmental differences. In addition, the size of maturity demonstrates plasticity even at the intra-population level (Duponchelle & Panfili, 1998).

Species reproductive strategy can indicate an allocation of energy resources to growth, with the objective of reaching the size at sexual maturity faster (Fontes-Filho, 2011). Against this background, knowledge on the $L_{50}$ of *P. managuensis*, as well as other attributes already mentioned, can contribute to future studies that propose to analyze possible advantages this parameter can offer about native species in the region. Besides the already mentioned characteristics, the identified length pattern ($L_{50}$) demonstrates that this possibly occurs due to changes in patterns of individual energy investment, reflected on the weight-length relationship (Freitas et al., 2016).

From the methods used to determine length at first maturity, the one by Freitas et al. (2016) shows the effective beginning of sexual maturity, as it portrays physiological oscillations resulting from patterns of energy investment per length phase. Therefore, it is the most recommended and the length that marks the beginning of maturity corresponds to 130.88 mm. The estimated number of cohorts for the species in this study was similar to that found by Farid et al. (2000) for the *P. dovii* population in Costa Rica: 3 cohorts. In that case, it should be noted that *P. managuensis* would not reach more than 2 years because of the quick growth already shown, adequate to the number of cohorts identified, with a six-monthly periodicity. This biological aspect agrees with the assumption that species with fast life cycles tend to reach a small size and exhibit low longevity (Fontes-Filho, 2011).

The observed length variations for the $L_{50}$ between *P. managuensis* and other species (Table 2) results from particularities of abiotic factors and interaction-derived pressures (Werner & Gilliam, 1984; Jones, 1992). For these reasons, when comparing the growth of *P. managuensis* in the present study to other cichlid species, it is noted that the observed variation arises from divergences between maximum sizes observed (Table 2). When analysing $K$ values for *P. managuensis* and the ones estimated for different cichlid species, it is possible to note that non-native species tend to exhibit a faster growth when compared to native species (Table 2). According to Sparre & Venema (1998), such differences are acceptable, since growth parameters can vary between species, in the same way as between springs for the same species. Still according to the authors, the growth coefficient is specifically related to metabolic rates, and vary

### Table 2. Maximum length ($TL_{max}$), von Bertalanffy growth parameters ($L_{\infty}$, $k$ and $t_0$) in growth studies for cichlid species estimated through length frequency distributions.

| Author             | Species                | $TL_{max}$ (mm) | $L_{\infty}$ (mm) | $k$ (ano$^{-1}$) | $t_0$ (years) | Country         | Native |
|-------------------|------------------------|-----------------|------------------|------------------|--------------|-----------------|--------|
| Jiménez-Badillo   | Oreochromis aureus      | 305.0           | 444.0            | 0.74             | 0.01         | Mexico          | No     |
| Jiménez-Badillo   | Cichla Piquiti         | 543.0           | 572.0            | 0.40             | -0.34        | Brazil          | No     |
| Gomiero et al.     | C. kelberi             | 548.0           | 577.0            | 0.35             | -0.39        | Brazil          | No     |
| Gomiero et al.     | C. kelberi             | 422.0           | 445.0            | 0.61             | -0.23        | Brazil          | No     |
| Campos & Freitas  | C. monoculus           | 510.0           | 535.5            | 0.38             | -            | Brazil          | Yes    |
| King & Etim       | Tilapia mariae         | 284.5           | 304.0            | 0.40             | -            | Nigeria         | Yes    |
| Present study     | Parachromis managuensis| 240.0           | 341.4            | 0.84             | -0.10        | Brazil          | No     |

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according to temperature and species biological factors, contributing to the present variations.

For the age and growth curve of \textit{P. managuensis} at the Saco I dam, it was noticed that the highest growth occurs until the first year of age. Thus, according to Pannella (1974), the decrease in growth rate in subsequent years occurs for tropical fishes, after first maturity, since a large part of the energy is targeted to reproduction. A fast growth pattern was noticed based on the analysis of the growth parameters for the studied species. In particular, the growth rate observed for \textit{P. managuensis} was certainly influenced by the characteristic climate of the semiarid, with a high evaporation rate playing a significant role in the organization and function of aquatic ecosystems, reflecting high temperature rates (Chellappa et al., 2009). This factor directly affects fish metabolism and growth, since the average temperature reported for the study site is 27.29 °C. Eventually, the growth rate for \textit{P. managuensis} in its natural habitat would presumably be different, since the climate representative to Costa Rica is milder, with an average annual temperature of 24.15 °C and a more pronounced rainy season (CRRH, 2008), factors that contribute to a slow growth (Jones, 1992; Sparre & Venema, 1997).

In general, natural mortality rates for tropical fishes like \textit{P. managuensis} tend to be higher due to multiple factors, such as predation and diseases that affect the population in different ways according to age (Pauly, 1998; Sparre & Venema, 1997). Mortality parameters for \textit{P. managuensis} were higher than those established for other cichlid species (Ruffino & Isaac, 1995; Mahmoud et al., 2013), even though it is considered as resistant and fast-adapting to abiotic factors such as oxygen concentration and pH fluctuations. Furthermore, it also exhibits parental care, which contributes to reduce its mortality (Berghe & Mckaye, 2001; Agasen et al., 2006; Rosana et al., 2006).

Based on the biological information obtained in this study and depletion of the native species \textit{Hoplias malabaricus} based on local fisher’s reports, we relate these factors to the presence of \textit{P. managuensis}. Its abundance in the dam has increased over the last 6 years, while \textit{H. malabaricus} disappeared from the dam within three years, and \textit{P. managuensis} is now fished and consumed in place of the tilapia (\textit{O. niloticus}), with a marked fisheries decline. The presence of \textit{P. managuensis} in the Saco I dam represents a risk to the native population of the trahira (\textit{Hoplias malabaricus}), hindering its growth. Since the introduced species has high resistance to abiotic oscillations and has eating habits similar to those of the native species, competing for food and space. The \textit{H. malabaricus} species in these conditions has a tendency to lose space due to its slow growth, which favors the dissemination of the introduced species (Barbieri, 1989; Martins, Rego, Pinese, 2009; Novaes & Carvalho, 2011). Under these conditions, \textit{P. managuensis} is also considered a real threat to fish fauna in the Rio Doce Basin, in Minas Gerais (Barros et al., 2012).

In summary, \textit{P. managuensis} showed a tendency for fast growth. However, it is advised that more precise studies must be carried out using hard structures (otolith) to better adjust the growth parameters, since in these structures it is possible to evaluate the real growth and age of the species since the larval stage, from the analysis of the growth rings. In addition, more studies aiming to improve knowledge on the biology of the species together with its mortality parameters in order to contribute towards the knowledge on possible impacts caused by \textit{P. managuensis} as an introduced species in Brazil.

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Ethical statement

The care and use of experimental animals are in compliance with [Brazilian Institute of the Environment and Renewable Natural Resources-IBAMA] laws, guidelines and animal welfare policies approved by [Chico Mendes Institute for Biodiversity Conservation-ICMBio, with license reference number 24709-1].

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