How alkaline compounds control atmospheric aerosol acidity

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Abstract. The acidity of atmospheric aerosols regulates the particulate mass, composition and toxicity, and has important consequences for public health, ecosystems and climate. Despite these broad impacts, the global distribution and evolution of aerosol acidity are unknown. We used the particular, comprehensive atmospheric multiphase chemistry – climate model EMAC to investigate the main factors that control aerosol acidity, and uncovered remarkable variability and unexpected trends during the past 50 years in different parts of the world. We find that alkaline compounds, notably ammonium, and to a lesser extent crustal cations, buffer the aerosol pH on a global scale. Given the importance of aerosols for the atmospheric energy budget, cloud formation, pollutant deposition and public health, alkaline species hold the key to control strategies for air quality and climate change.

1. Introduction

Aerosol acidity is a central property of atmospheric particulates that influence clouds, climate and air quality, including impacts on human health (Raizenne et al., 1996; Lelieveld et al., 2015). It affects the partitioning of semi-volatile acids between the gas and aerosol phases (Guo et al., 2016; Guo et al., 2017; Guo et al., 2018; Nenes et al., 2020), secondary organic aerosol (SOA) formation (Xu et al., 2015; Marais et al., 2016), the solubility of trace metals in aerosols (Oakes et al., 2012), associated with their toxicity (Fang et al., 2012; Fang et al., 2017) and nutrient capacity (Jickells et al., 2005), the activation of halogens that act as oxidants (Saiz-Lopez and von Glasow, 2012), the conversion of sulfur dioxide (Seinfeld and Pandis, 2006; Cheng et al., 2016), the particle hygroscopic growth and lifetime (Metzger et al., 2006; Abdelkader et al., 2015; Karydis et al., 2016, 2017), and atmospheric corrosivity (Leygraf et al., 2016). Direct measurement of aerosol acidity is difficult and associated with much uncertainty, being dependent on filter sampling and the H⁺ molality in the aqueous extract, which is sensitive to artifacts (Pathak et al., 2004; Pathak et al., 2004). Therefore, particle
pH, a commonly used acidity metric of aqueous aerosols, is typically inferred by proxy techniques (Hennigan et al., 2015; Pye et al., 2020). Two of the most common are the ion balance and the molar ratio methods. In the past, these methods did not consider the effects of aerosol water and multiphase interactions with gas phase species as well as the partial dissociation of acids (Hennigan et al., 2015). The simultaneous measurement of gas phase species can improve aerosol pH estimates by accounting for the phase partitioning of semi-volatile species (e.g., NH₃, HNO₃). However, the accuracy of this approach relies on the availability of information on these species in both the gas and aerosol phase, being scant in most cases.

The best estimates of pH are obtained with thermodynamic equilibrium models, although the accuracy can be limited by not accounting for all ionic species. For example, most atmospheric chemistry models do not consider crustal elements (e.g., Ca²⁺, Mg²⁺, K⁺, Na⁺) and Na⁺ in sea salt. These species affect the ion balance by influencing the phase partitioning of nitrate and ammonium, especially in areas where aeolian dust is abundant (Karydis et al., 2016). Here we present 50-year global acidity trends of fine aerosols (i.e., with a diameter < 2.5 μm) by employing the EMAC chemistry–climate model (Jöckel et al., 2010). The pH calculations are performed online with the ISORROPIA II thermodynamic equilibrium model (Fountoukis and Nenes, 2007).

2. Results and Discussion

2.1 Global variability of aerosol acidity

Figure 1 shows the modeled near-surface distribution of fine aerosol acidity for the 2010-2015 period. We find predominantly acidic particles over the anthropogenically influenced regions in the northern hemisphere and the tropical biomass burning zones, and mostly alkaline particles over deserts and oceans, especially over the southern oceans. The pH typically ranges from 4.0 to 6.7 (5.3 on average) over the western USA since it is affected by crustal cations from the surrounding deserts. Therefore, polluted areas located downwind of crustal sources are of special interest since the pH calculations in this region are sensitive to the aerosol state assumption (see section 4.3). Over Pasadena, the base case model using the stable state mode estimates a mean pH of 5.9 units, while the sensitivity simulation with only liquid aerosol results in 2.7 pH units (equal to Guo et al. (2017) estimations by using the metastable assumption; Table S1). Our sensitivity analysis revealed that the aerosol state itself is not affected by the state assumption since both stable and metastable predict the same amount of water in the aerosol. Differences in the calculated pH can be due to the high concentrations of calcium from the Great Basin Desert which results in the precipitation of high amounts of CaSO₄, lowering the particle acidity (but without affecting the water activity since CaSO₄ is insoluble and does not contribute to the MDRH depression). It is worth mentioning that calcium was not included in the Guo et al. (2017) study which helps explain the differences in the observed and simulated aerosol acidity. The simulated particle-phase fraction of nitrate over Pasadena is 40% using the stable state assumption and 32% using the metastable assumption, compared to the observationally derived 51%. Over Europe, the pH
ranges from 2.6 to 6.7 (3.9 on average). Observational estimates of aerosol pH from the Po Valley (Squizzato et al., 2013; Masiol et al., 2020) and Cabauw (Guo et al., 2018) support the relatively low acidity of fine aerosols over Europe (Table S1). Model calculations compare well with observational estimates from Cabauw, however, result in higher pH (~1 unit) compared to values from Po Valley (estimated by using the E-AIM model). Over East Asia the average pH is 4.7, ranging from 2.6 to 7.4. Relatively high pH are found over regions where anthropogenic aerosols are mixed with aeolian dust, e.g., from the Gobi Desert, which buffer decrease the acidity (e.g., ~6 pH units over Hohhot, which agrees well with the estimations of Wang et al. (2019a)). The relatively low pH in large parts of Asia is explained by strong SO$_2$ emissions and associated sulfate, which have increased strongly in the past decades (e.g., over Guangzhou, supported by estimations of Jia et al. (2018)). Estimates of unrealistically high aerosol acidity can result from omitting the gas phase concentrations of semi-volatile ions from the pH calculations (e.g., estimates over Hong Kong (Yao et al., 2007; Xue et al., 2011), Singapore (Behera et al., 2013) and Shanghai (Pathak et al., 2009); Table S1). At the same time, SO$_2$ emissions have decreased over Europe and USA, and recently in China. However, aerosols over the eastern USA have remained acidic, with an average pH of 3.0 until recently, corroborating the findings of Weber et al. (2016) and Lawal et al. (2018) that aerosol acidity over this region is less sensitive to SO$_2$ than to NH$_3$ emissions.

The aerosol pH over the anthropogenically influenced/polluted northern hemispheric mid-latitudes (e.g., over East Asia) and the northern extratropical oceans exhibits a clear seasonal pattern with lower values during boreal summer and higher ones during winter, driven by the availability of ammonium and by the aerosol water content (Fig. 2). This is evident from both our model calculations and from observational estimates mostly in heavily populated areas such as the Po Valley (Squizzato et al., 2013), Beijing (Tan et al., 2018), and Tianjin (Shi et al., 2017), and to a lesser extent over areas strongly affected by aeolian dust (e.g., Hohhot; Wang et al., 2019b) (Table S1). Over tropical regions, fine particulates have a pH between 3.2 and 7.4, being strongly influenced by pyrogenic potassium, i.e., from widespread biomass burning (Metzger et al., 2006), and a high aerosol water content. Observational estimates from Sao Paulo support these high pH values (Vieira-Filho et al., 2016), albeit with 1 unit bias mainly related to the use of the E-AIM model. Over deserts, aerosols are relatively alkaline, with a pH up to 7.4. Aerosols in the marine environment tend to be alkaline also, with a pH up to 7.4 over the southern oceans. Observational estimates report highly acidic aerosols over the southern oceans due to the lack of gas phase input for the pH calculations (Dall’Osto et al., 2019). Over the Arctic and the northern Atlantic and Pacific Oceans, aerosol acidity is significantly enhanced by strong sulfur emissions from international shipping and pollution transport from industrialized areas (Fig. 1). The pH over the northern extratropical oceans and the Arctic ranges from 2.0 to 7.0 with an average of about 5.2. The annual cycle of aerosol acidity over these regions is strongly influenced by anthropogenic pollution, being relatively high during boreal summer. Over the Antarctic, aerosol pH ranges from 4.5 to 7.0 and follows a clear seasonal pattern (Fig. 2).
2.2 Temporal evolution of aerosol acidity

Figure 1 and Table 1 present the aerosol pH over the period 1970-2020. We investigated the impacts of alkaline species by omitting the emissions of ammonia and mineral cations in two sensitivity simulations.

2.2.1 Europe

Over Europe, the pH has increased strongly from about 2.8 during the 1970s to 3.9 recently. Especially during the 1990s NH$_3$ emissions over Europe increased significantly by 14%, while at the same time NOx and SO$_2$ emissions decreased by 13% and 49%, respectively. While this trend has continued in the past decade, pH changes slowed because the sulfate and nitrate decreases have been compensated through volatilization of ammonia from the particles. In addition, the recently increasing cation/anion ratio is accompanied by a reduction of aerosol water, preventing a significant decrease of the aerosol acidity (Fig. S1). Overall, the increase of aerosol pH by more than 1 unit during the last 50 years had a significant impact on the gas-particle partitioning of semi-volatile acids, e.g., nitric acid, since their dissociation into ions enhances their solubility (Nah et al., 2018). Here, the fraction of nitrate in the particle phase relative to total nitrate (gas plus particle) has increased from ~70% to 85% (Fig. 3). The increase in aerosol pH has been accompanied by an increase in aerosol hygroscopicity (Fig. 4). After the substantial reduction of SO$_2$ emissions, sulfate salts (e.g., ammonium sulfate with $\kappa = 0.53$) are replaced by more hygroscopic nitrate salts (e.g., ammonium nitrate with $\kappa = 0.67$) in the aerosol composition. In addition, the decrease of organic compound emissions during the last 50 years contributed to the increase of the aerosol hygroscopicity. Our sensitivity simulations reveal that aerosol acidity over Europe is highly sensitive to NH$_3$ emissions. Despite the decline of both SO$_2$ and NOx during the past decades, the aerosol would have remained highly acidic (pH ~1) in the absence of NH$_3$.

2.2.2 North America

Over North America, aerosol acidity also decreased with SO$_2$ and NOx emissions. Nevertheless, these emissions are still relatively strong in the eastern USA (5 times higher than in the western USA) resulting in very acidic aerosols, with a pH ranging from 2.2 in 1971 to 3.3 recently (Figs. 1 and S1). Such acidic conditions promote the dissolution of metals (e.g., Fe, Mn, Cu) in ambient particles (Fang et al., 2017; Fang et al., 2017). Soluble transition metals in atmospheric aerosols have been linked to adverse health impacts since they generate reactive oxygen species, leading to oxidative stress and increased toxicity of fine particulate matter (Fang et al., 2017; Park et al., 2018). Since the solubility of transition metals increases exponentially below a pH of 3, the decrease of aerosol acidity over the eastern USA reported here suggests that the particles have become substantially less toxic in the past few decades. Similar to Europe, the increasing pH has resulted in a growing aerosol nitrate fraction from ~50% during the 1970s to 65% recently (Fig. 3), and to a strong increase of aerosol hygroscopicity by ~0.15 units at the cloud base (Fig. 4). The role of NH$_3$ is critically important; without it the aerosol pH
over the eastern USA would be close to zero. Over the western USA, the aerosol pH is higher (~5), being affected by aeolian dust from the Great Basin Desert, although NH$_3$ is still the most important alkaline buffer.

### 2.2.3 East and South Asia

In Asia, SO$_2$ and NOx emissions have increased drastically since 1970. However, the simultaneous increase of NH$_3$ emissions along with the presence of mineral dust from the surrounding deserts (i.e., Gobi, Taklimakan, Thar) decelerated the increase of aerosol acidity. Over East Asia, the aerosol pH decreased from about 5.3 during the 1970s to 4.5 in 2010. This change in aerosol acidity has affected the predominant pathway of sulfate formation through aqueous phase chemistry. Under acidic conditions, SO$_2$ is mainly oxidized by dissolved H$_2$O$_2$, while at pH > 5 the oxidation by O$_3$ predominates (Seinfeld and Pandis, 2006). Aerosol aqueous phase chemistry. Under acidic conditions, SO$_2$ is mainly oxidized by transition metal ions, while at pH > 5 the oxidation by O$_3$ and NO$_2$ predominates (Cheng et al., 2016). Therefore, the decrease of pH during the last 50 years, even though being relatively modest, was sufficient to turn-off sulfate production from O$_3$ oxidation (Fig. 5). At the same time, the increased aerosol acidity hinders the partitioning of nitric acid to the aerosol phase, reducing the aerosol nitrate fraction from 90% to 80% (Fig. 3). Remarkably, the aerosol hygroscopicity has increased from ~0.3 in the 1970s to 0.45 recently (Fig. 4), revealing a reverse development compared to Europe and the USA. Here, the fraction of mineral dust in the aerosol is higher; therefore, the particles gained hygroscopicity by the acquired pollution solutes. Recently, the SO$_2$ emissions have dropped and the NOx emission increase has slowed in East Asia, while SO$_2$ emissions are soaring in South Asia. SO$_2$ emission trends since 2007 have been so drastic that inventories and scenarios tend to underestimate them. Satellite observations indicate that India has recently overtaken China as the world largest emitter of SO$_2$ (Li et al., 2017). Following the satellite observations, we implemented the large significant SO$_2$ reduction trends into our model (Fig. S2). Surprisingly, the effect only becomes noticeable over East Asia after 2016, when the aerosol pH started increasing by about 0.3 units, while we do not find any change over South Asia. This corroborates the strong buffering that we found over other regions such as Europe. Fig. 1 shows that NH$_3$ has been the major buffer, supporting the recent findings of Zheng et al. (2020) that the acid-base pair of NH$_3$/NH$_4^+$ provides the largest buffering capacity over East and South Asia. However, we also found that in East Asia and to a lesser extent in South Asia, crustal elements, not considered in the study of Zheng et al. (2020), have contributed significantly on maintaining a mean pH of 4.5–5 in the past decade (Fig. 1). Calcium is the major crustal component of dust from the Gobi and Taklimakan deserts (Karydis et al., 2016) and unlike other crustal compounds it can react with sulfate ions and form insoluble CaSO$_4$, which precipitates out of the aerosol aqueous phase. This interaction reduces the aqueous sulfate and thus the aerosol acidity.

### 2.2.4 Tropical forests, Middle East

Over tropical forests, aerosols are typically not very acidic with pH values >4. Note that organic acids were not included in the aerosol pH calculations, however, their contribution to the total ionic load is small (Andreae et al., 1988; Falkovich et al., 2005), and aerosol acidity can be attributed to inorganic acids. Over the Amazon and Congo basins, the aerosol pH remained
around 5 since 1970. The Southeast Asian forest atmosphere is affected by pollution from mainland Asia, and the aerosol pH
decreased to around 4 recently. This pH drop has enhanced SOA formation from isoprene, since under low-NOx conditions
typical over rainforests) the presence of acidifying sulfate increases the reactive uptake of epoxydiols (Xu et al.,
2015; Surratt et al., 2010). Nevertheless, NH$_3$ emissions provide a remarkably strong buffer over all three tropical regions
while mineral dust cations are also important over the Amazon and Congo forests. Further, the Middle East is affected by
strong anthropogenic (fossil fuel related) and natural (aeolian dust) aerosol sources. Due to the high abundance of mineral
dust, the pH has remained close to 7. Without crustal cations, the pH would drop to about 4. Despite the omnipresence of
alkaline species from the surrounding deserts, NH$_3$ still plays a central role in controlling the acidification of mineral dust
aerosols, which can affect their hygroscopic growth and hence their climate forcing (Klingmuller et al., 2019; Klingmüller et
al., 2020).

2.2.5 Oceans

Over the Arctic and northern extra-tropical oceans, aerosol acidity is strongly affected by pollution transport from the urban-
industrial mid-latitudes. The Arctic aerosol pH is highly variable, remaining relatively low up to 1990 (~4.2), after which it
increased to about 5.2. Crustal cations are found to play a significant buffering role lowering the aerosol acidity. Over the
northern extra-tropical oceans, aerosol pH has remained relatively constant (~4.8). NH$_3$ provides an important alkaline
buffer, and without it the aerosol pH would have been below 3. NH$_3$ is also proved to be important over the tropical and
southern extra-tropical oceans, where a noticeable increase in aerosol acidity occurred after June 1991, when the eruption of
Mount Pinatubo in the Philippines released ~20 million tons of SO$_2$ into the stratosphere (McCormick et al., 1995). The
impact of Pinatubo sulfate, after returning to the troposphere, on aerosol acidity is mostly evident over Antarctica, where the
pH dropped by 2 units, as the stratospheric circulation is strongest in the winter hemisphere. Over Antarctica concentrations
of dust and especially of NH$_3$ are very low, and Fig. 1 illustrates that only in this pristine environment the large Pinatubo
anomaly could overwhelm the buffering by alkaline species. Except after Pinatubo, the pH has remained nearly constant at
5.8 over Antarctica and about 5.5 in the tropics and 6.8 in the southern extra-tropics.

3. Conclusions

We find that aerosol pH is generally well-buffered by alkaline compounds, notably NH$_3$ and in some areas crustal elements.
NH$_3$ is found to supply remarkable buffering capacity on a global scale, from the polluted continents to the remote oceans. In
the absence of NH$_3$, aerosols would be highly (to extremely) acidic in most of the world. Therefore, potential future changes
in NH$_3$ are critically important in this respect. Agriculture is the main NH$_3$ source and a controlling factor in fine particle
concentrations and health impacts in some areas (e.g., Europe) (Pozzer et al., 2017). The control of agricultural ammonia
emissions must therefore be accompanied by very strong reductions of SO$_2$ and NOx to avoid that aerosols become highly
acidic with implications for human health (aerosol toxicity), ecosystems (acid deposition and nutrient availability), clouds
and climate (aerosol hygroscopicity).

4. Appendix A: Materials and Methods

4.1 Aerosol-chemistry-climate model

We used the ECHAM5/MESSy Atmospheric Chemistry (EMAC) model, which is a numerical chemistry and climate simulation system that describes lower and middle atmosphere processes (Jöckel et al., 2006). EMAC uses the Modular Earth Submodel System (MESSy2) (Jöckel et al., 2010) to link the different sub-models with an atmospheric dynamical core, being an updated version of the 5th generation European Centre - Hamburg general circulation model (ECHAM5) (Roeckner et al., 2006). EMAC has been extensively described and evaluated against in situ observations and satellite retrievals to compute particulate matter concentrations and composition, aerosol optical depth, acid deposition, gas phase mixing ratios, cloud properties, and meteorological parameters (Karydis et al., 2016; Pozzer et al., 2012; Tsimpidi et al., 2016; Karydis et al., 2017; Bacer et al., 2018). The spectral resolution of EMAC used in this study is T63L31, corresponding to a horizontal grid resolution of approximately 1.9°x1.9° and 31 vertical layers extending up to 10 hPa (i.e., 25 km) from the surface. The presented model simulations encompass the 50-year period 1970-2020.

EMAC calculates fields of gas phase species online through the Module Efficiently Calculating the Chemistry of the Atmosphere (MECCA) Submodel (Sander et al., 2019). MECCA calculates the concentration of a range of gases, including aerosol precursor species (e.g. SO$_2$, NH$_3$, NO$_x$, DMS, H$_2$SO$_4$ and DMSO) and the major oxidant species (e.g. OH, H$_2$O$_2$, NO$_3$, and O$_3$). Aerosol microphysics are calculated by the Global Modal-aerosol eXtension (GMXe) module (Pringle et al., 2010). The organic aerosol formation and atmospheric evolution are calculated by the ORACLE Submodel (Tsimpidi et al., 2014, 2018). The aerosol size distribution is described by seven lognormal modes: four hydrophilic modes that cover the same size range except nucleation. The aerosol composition within each size mode is uniform (internally mixed), however, it varies between modes (externally mixed). Each mode is defined in terms of total number concentration, number mean radius, and geometric standard deviation (Pringle et al., 2010). The removal of gas and aerosol species through wet and dry deposition is calculated within the SCAV (Tost et al., 2006) and DRYDEP (Kerkweg et al., 2006) submodels, respectively. The sedimentation of aerosols is calculated within the SEDI submodel (Kerkweg et al., 2006). The cloud cover, microphysics and precipitation of large scale clouds is calculated by the CLOUD Submodel (Roeckner et al., 2006) which uses a two-moment stratiform microphysical scheme (Lohmann and Ferrachat, 2010), and describes liquid droplet (Karydis et al., 2017) and ice crystal (Bacer et al., 2018) formation by accounting for the aerosol physicochemical properties. The effective hygroscopicity parameter $\kappa$ is used to describe the influence of chemical composition on the cloud condensation nuclei (CCN) activity of atmospheric aerosols. $\kappa$ is calculated using the mixing rule of Petters and Kreidenweis (Petters and
and the individual $\kappa$ parameter values for each inorganic salt (Petters and Kreidenweis, 2007; Sullivan et al., 2009). Organic aerosol species are assumed to have a constant hygroscopicity $\kappa$ parameter of 0.14 while bulk mineral dust and black carbon are assumed to have zero hygroscopicity.

### 4.2 Emissions

The vertically distributed (Pozzer et al., 2009) CMIP5 RCP8.5 emission inventory (van Vuuren et al., 2011) is used for the anthropogenic and biomass burning emissions during the years 1970-2020. Direct emissions of aerosol components from biofuel and open biomass burning are considered by using scaling factors applied on the emitted black carbon based on the findings of Akagi et al. (Akagi et al., 2011) (Table S2). Dust emission fluxes and emissions of crustal species ($\text{Ca}^{2+}, \text{Mg}^{2+}, \text{K}^+, \text{Na}^+$) are calculated online as described by Klingmüller, et al. (Klingmüller et al., 2018) and based on the chemical composition of the emitted soil particles in every grid cell (Karydis et al., 2016); Table S3. NO$_x$ produced by lightning is calculated online and distributed vertically based on the parameterization of Grewe, et al. (Grewe et al., 2001). The emissions of NO from soils are calculated online based on the algorithm of Yienger and Levy (Yienger and Levy, 1995). The oceanic DMS emissions are calculated online by the AIRSEA Submodel (Pozzer et al., 2006). The natural emissions of NH$_3$ are based on the GEIA database (Bouwman et al., 1997). Emissions of sea spray aerosols (assuming a composition suggested by Seinfeld and Pandis (Seinfeld and Pandis, 2006); Table S2) and volcanic degassing emissions of SO$_2$ are based on the offline emission data set of AEROCOM (Dentener et al., 2006).

### 4.3 Thermodynamic model

The inorganic aerosol composition, which is of prime importance for the accurate pH calculation, is computed with the ISORROPIA-II thermodynamic equilibrium model (Fountoukis and Nenes, 2007). ISORROPIA-II calculates the gas/liquid/solid equilibrium partitioning of the $\text{K}^+\text{-Ca}^{2+}\text{-Mg}^{2+}\text{-NH}_4^+\text{-Na}^+\text{-SO}_4^{2-}\text{-NO}_3^\text{-Cl}^{-}\text{-H}_2\text{O}$ aerosol system and considers the presence of 15 aqueous phase components and 19 salts in the solid phase. ISORROPIA-II solves for the equilibrium state by considering the chemical potential of the species and minimizes the number of equations and iterations required by considering specific compositional “regimes”. Furthermore, the assumption of thermodynamic equilibrium is a good approximation for fine-mode aerosols that rapidly reach equilibrium. However, the equilibrium timescale for large particles is typically larger than the time step of the model (Meng and Seinfeld, 1996) leading to errors in the size distribution of semi-volatile ions like nitrate. Since the current study include reactions of nitric acid with coarse sea-salt and dust aerosol cations, the competition of fine and coarse particles for the available nitric acid can only be accurately represented by taking into account the kinetic limitations during condensation of HNO$_3$ in the coarse mode aerosols. To account for kinetic limitations by mass transfer and transport between the gas and particle phases, the process of gas/aerosol partitioning is calculated in two stages (Pringle et al., 2010). First, the gaseous species that kinetically condense onto the aerosol phase...
within the model timestep are calculated assuming diffusion limited condensation (Vignati et al., 2004). Then, ISORROPIA-II re-distributes the mass between the gas and the aerosol phase assuming instant equilibrium between the two phases.

ISORROPIA-II is used in the forward mode, in which the total (i.e., gas and aerosol) concentrations are given as input. Reverse mode calculations (i.e. when only the aerosol phase composition is known) should be avoided since they are sensitive to errors and infer bimodal behavior with highly acidic or highly alkaline particles, depending on whether anions or cations are in excess (Song et al., 2018). While it is often assumed that aerosols are in a metastable state (i.e., composed only of a supersaturated aqueous phase), here we use ISORROPIA-II in the thermodynamically stable state mode where salts are allowed to precipitate once the aqueous phase becomes saturated. For this purpose, we have used the revised ISORROPIA-II model which includes modifications proposed by Song et al. (2018), who resolved coding errors related to pH calculations when the stable state assumption is used. A sensitivity simulation with only liquid aerosols (i.e., metastable) revealed that the assumed particle phase state does not significantly impact the pH calculations over oceans and polluted regions (e.g., Europe), however, the metastable assumption produces more acidic particles (up to 2 units of pH) in regions affected by high concentrations of crustal cations (Fig. S3). Overall, the stable state assumption used here produces about 0.5 units higher global average pH than the metastable assumption. By comparing with the benchmark thermodynamic model E-AIM, Song et al. (2018) found that ISORROPIA-II produces somewhat higher pH (by 0.1-0.7 units, negatively correlated with RH). However, E-AIM model versions either lack crustal cations from the ambient mixture of components (e.g. version II) (Clegg et al., 1998), or only include Na+ with the restriction that it should be used when RH > 60% (e.g. version IV) (Friese and Ebel, 2010). - Song et al. (2018) applied the revised ISORROPIA-II during winter haze events in eastern China and found that the assumed particle phase state, either stable or metastable, does not significantly impact the pH predictions.

We performed a sensitivity simulation with only liquid aerosols (i.e., metastable), which revealed that the assumed particle phase state does not significantly impact the pH calculations over oceans and polluted regions (e.g., Europe), however, the metastable assumption produces more acidic particles (up to 2 units of pH) in regions affected by high concentrations of crustal cations and consistently low RH values (Fig. S3). Fountoukis et al. (2007) have shown that the metastable solution predicts significant amounts of water below the mutual deliquescence relative humidity (MDRH, where all salts are simultaneously saturated with respect to all components). Further, the generally high calcium concentrations downwind of deserts results in increasing pH values due to the precipitation of insoluble salts such as the CaSO4. The metastable state assumption fails to reproduce this since it treats only the ions in the aqueous phase. In general, high amounts of crustal species can significantly increase the aerosol pH which is consistent with the presence of excess carbonate in the aerosol phase (Meng et al., 1995). It is worth mentioning that the stable state solution algorithm of ISORROPIA II starts with assuming a dry aerosol, and based on the ambient RH dissolves each of the salts depending on their DRH. However, in the ambient atmosphere, when the RH over a wet particle is decreasing, the wet aerosol may not crystallize below the MDRH but instead remain in a metastable state affecting the uptake of water by the aerosol and thus the pH. This could be the case in some locations with high diurnal variations of RH. Our sensitivity calculations show that, overall, the stable state assumption produces an about 0.5 units higher global average pH than the metastable assumption. Karydis et al. (2016) have
shown that while the aerosol state assumption has a marginal effect on the calculated nitrate aerosol tropospheric burden (2% change), it can be important over and downwind of deserts at very low RHs where nitrate is reduced by up to 60% by using the metastable assumption. This is in accord with the findings of Ansari and Pandis (2000) who suggested that the stable state results in higher concentrations of aerosol nitrate when the RH is low (<35 %) and/or sulfate to nitrate molar ratios are low (<0.25).

4.4 pH calculations

The pH is defined as the negative decimal logarithm of the hydrogen ion activity \( a_{\text{H}^+} = \gamma x_{\text{H}^+} \) in a solution:

\[
\text{pH} = -\log_{10}(x_{\text{H}^+}) \]

where \( x_{\text{H}^+} \) is the molality of hydrogen ions in the solution and \( \gamma \) is the ion activity coefficient of hydrogen. Assuming that \( \gamma \) is unity, the aerosol pH can be calculated by using the hydrogen ion concentration in the aqueous aerosol phase calculated by ISORROPIA-II (in mole m\(^{-3}\)) and the aerosol water content calculated by GMXe (in mole Kg\(^{-1}\)). GMXe assumes that particle modes are internally mixed, and takes into account the contribution of both inorganic and organic (based on the organic hygroscopicity parameter, \( \kappa_{\text{org}} = 0.14 \) (Tsimpidi et al., 2014)) species to aerosol water.

The aerosol pH is calculated online at each timestep, and output stored every five hours based on instantaneous concentrations of fine aerosol water and hydrogen ions. The average pH values shown in the manuscript are based on the calculated instantaneous mean pH values. According to the Jensen’s inequality (Jensen, 1906), the average of the instantaneous pH values is less than or equal to the pH calculated based on the average of the water and hydrogen ion instantaneous values. We estimate that the average pH calculated based on 5-hourly instantaneous values is approximately 1-3 (-2 globally averaged) units higher than the pH calculated based on the average water and hydrogen ion concentrations. By including online gas-particle partitioning calculations of the NH\(_3\)/HNO\(_3\) system in polluted air, as applied here, we find that the aerosol pH is higher by approximately one unit (Guo et al., 2015). Hence by neglecting these aspects the aerosol pH would be low-biased by about 3 units.

4.5 Comparison against pH estimations from field derived PM\(_{2.5}\) compositional data

The pH calculated here is compared against pH estimations from field derived PM\(_{2.5}\) compositional data around the world compiled by Pye et al. (2020) (Table S1). pH data derived from other aerosol sizes (e.g., PM\(_{1}\)) has been omitted since aerosol acidity can vary significantly with size (Zakoura et al., 2020). It should be emphasized that the comparison presented
in Table S1 aims to corroborate the spatial variability of pH found in this study and not to strictly evaluate the model calculations. Observationally estimated aerosol pH is derived from a variety of methods that can affect the result significantly as discussed above (i.e., the use of E-AIM or ISORROPIA, stable/metastable assumption, forward/reverse mode, and the availability of gas phase NH$_3$/HNO$_3$, crustal species, and organic aerosol water observations).—evaluate the model calculations. Since direct measurements of aerosol acidity are not available, the observation-based aerosol pH is estimated by employing thermodynamic equilibrium models (e.g., ISORROPIA) and making assumptions that can significantly affect the results, especially when the data are averaged over extended periods, while RH conditions during data collection are not always accounted for, e.g. in studies based on filter sampling. The calculation of aerosol acidity on a global scale requires the advanced treatment of atmospheric aerosol chemical complexity, representing the real atmosphere, and beyond the conventional methods used by chemistry-climate models (CCM). The atmospheric chemistry model system EMAC is an ideal tool for this purpose since it is one of the most comprehensive CCM containing advanced descriptions of the aerosol thermodynamics (including e.g. dust-pollution interactions) and organic aerosol formation and atmospheric aging (affecting the aerosol water). Our model calculations for aerosol acidity are based on some processes/factors that are not included explicitly, usually neglected by model calculations used to constrain the aerosol acidity from observations. Sources of discrepancy between the pH calculations can be the following:

- The stable/metastable assumption does not affect the pH most of the time, however, in some cases with low RHs and the presence of crustal cations, the metastable assumption results in lower pHs (see section 4.3).
- Crustal species from deserts and Na$^+$ from sea salt can elevate the pH significantly in some locations, however, these are often neglected in observations.
- The organic aerosols (which are treated comprehensively by our model using the module ORACLE and the volatility basis set framework (Tsimpidi et al., 2014)) can contribute significantly to the aerosol water, and thus increase the aerosol pH. This contribution is not considered by many observational studies.
- Including gas phase species (e.g., NH$_3$, HNO$_3$) in the pH calculations is important. Using only the aerosol-phase as input (i.e., reverse mode) the inferred pH exhibits bimodal behaviour with very acidic or alkaline values depending on whether anions or cations are in excess (Hennigan et al., 2015). Even if the forward mode is used (without gas phase input), the calculated aerosol pH is biased low (approximately 1 pH unit) due to the repartition of semi-volatile anions (i.e., NH$_3$) to the gas phase to establish equilibrium (Guo et al., 2015).
- Another important aspect, not explicitly mentioned in many studies, relates to the methods used to derive the campaign-average (or for 3D models the simulated average) pH. In our model the aerosol pH is calculated online (2-minute time resolution), while output is stored every five hours based on instantaneous concentrations of fine aerosol H$_2$O and H$^+$. This mimics 5-hourly aerosol sampling. Then, the average pH values are calculated from the instantaneous mean pH values (see section 4.4). Often models use average values (and not instantaneous) as output, or field-derived pH calculations use average observed H$_2$O and H$^+$ values, which can result in important underestimation (by ~ 1-3 units) of the aerosol pH (Jensen, 1906).
• Some unrealistically high pH values in a few past studies resulted from coding errors in the stable state assumption of the ISORROPIA II model, which have been corrected in our study following the recommendation of Song et al. (2018).

• The type of thermodynamic model used is also important. Song et al. (2018) found that ISORROPIA-II produces somewhat higher pH (by 0.1-0.7 units, negatively correlated with RH) compared to the thermodynamic model E-AIM, which is used to observationally-constrain pH in some studies.

• Measurements of PM2.5 nitrate are not always reliable because of artifacts associated with the volatility of ammonium nitrate (Schaap et al., 2004). Ammonium and nitrate can partially evaporate from Teflon filters at temperatures between 15 to 20 °C and can evaporate completely at temperatures above. The evaporation from quartz filters is also significant at temperatures higher than 20 °C. This systematic underestimation of ammonium nitrate can affect the observed chemical composition of the aerosol and thus the pH calculations.

• The comparison between global model output and observations at specific locations. This also concerns the aerosol concentrations but is especially important for the aerosol acidity. Apart from the size of the model grid cells (i.e., ~1.9’x1.9’), the altitude is also important. The first vertical layer of EMAC is approximately 67m in height. On the other hand, ground observations are typically collected in a height up to 3 m. While the aerosols within size modes simulated in our model are well-mixed, perhaps this is not the case for the aerosols observed at the surface and potentially close to sources, and thus the aerosol acidity may be higher (e.g., due to the higher contribution from local primary sources like SO4^2-, lower water amounts in the aerosol, or lower concentrations of semi-volatile cations like NH4^+).

4.6 Emissions

The vertically distributed (Pozzer et al., 2000) CMIP5 RCP8.5 emission inventory (van Vuuren et al., 2011) is used for the anthropogenic and biomass burning emissions during the years 1970-2020. Direct emissions of aerosol components from biofuel and open biomass burning are considered by using scaling factors applied on the emitted black carbon based on the findings of Akagi, et al. (Akagi et al., 2011) (Table S2). Dust emission fluxes and emissions of crustal species (Ca^{2+}, Mg^{2+}, K^+, Na^+) are calculated online as described by Klingmuller, et al. (Klingmuller et al., 2018) and based on the chemical composition of the emitted soil particles in every grid cell (Karydis et al., 2016; Table S3). NOx produced by lightning is calculated online and distributed vertically based on the parameterization of Grewe, et al. (Grewe et al., 2001). The emissions of NO from soils are calculated online based on the algorithm of Yienger and Levy (Yienger and Levy, 1995). The oceanic DMS emissions are calculated online by the AIRSEA Submodel (Pozzer et al., 2006). The natural emissions of NH3 are based on the GEIA database (Bouwman et al., 1997). Emissions of sea spray aerosol (assuming a composition suggested by Seinfeld and Pandis (Seinfeld and Pandis, 2006); Table S2) and volcanic degassing emissions of SO2 are based on the offline emission data set of AGROCOM (Dentener et al., 2006).
4.5 Partitioning of nitric acid between the gas and aerosol phases

The impact of pH on the fraction of nitrate in the particle phase relative to total nitrate (gas plus particle), i.e., $\varepsilon(\text{NO}_3^-)$, during the 50 years of simulation in specific regions is calculated as follows (Nah et al., 2018):

$$\varepsilon(\text{NO}_3^-) = \frac{H_{\text{HNO}_3}\text{WRT}(0.987 \times 10^{-14})}{\gamma_{\text{NO}_3^-}\gamma_H^+10^{-pH} + H_{\text{HNO}_3}\text{WRT}(0.987 \times 10^{-14})} \quad (A2)$$

Where $H_{\text{HNO}_3}$ is the combined molality-based equilibrium constant of HNO$_3$ dissolution and deprotonation, $\gamma$’s represent the activity coefficients, W is the aerosol water, R is the gas constant, and T is the ambient temperature. $\varepsilon$ of $\text{NO}_3^-$ is equivalent with the instantaneous calculations of ISOROPIA II within EMAC. However, the model output is produced after considering all processes in the model and is not calculated at every timestep. Therefore, the use of Eq. 2 can provide a clearer picture of the impact of pH on HNO$_3$ gas/particle partitioning since the model output (e.g., gas-phase HNO$_3$ and nitrate in 4 size modes) is subject to uncertainties related to other processes (e.g., deposition, coagulation, transport, etc.).

4.6 Sulfate formation in aqueous aerosols

The sulfate production rate on aqueous aerosols from the heterogeneous oxidation of S(IV) with the dissolved O$_3$ is given by

$$R_0 = k [\text{O}_3^-] \quad (A3)$$

The first-order uptake rate, $k$, from monodisperse aerosols with radius $r_a$ and total aerosol surface $A$, is calculated following Jacob (Jacob, 2000):

$$k = \left( \frac{r_a}{D_g} + \frac{4}{\nu \gamma} \right)^{-1} A \quad (A4)$$

where $\nu$ is the mean molecular speed of O$_3$ and $D_g$ is its gas-phase molecular diffusion coefficient calculated as follows:

$$D_g = \frac{9.45 \times 10^{17} \times T \left( 3.47 \times 10^{-2} + \frac{1}{M} \right)}{\rho_{\text{air}}} \quad (A5)$$

where $T$ is the ambient air temperature, $\rho_{\text{air}}$ is the air density, and $M$ the molar mass of O$_3$. $\gamma$ is the reaction probability calculated following Jacob (Jacob, 2000) and Shao et al. (Shao et al., 2019).

$$\gamma = \frac{1}{\alpha} + \frac{\nu}{4HRT D_a K f_r} \quad (A6)$$

where $\alpha$ is the mass accommodation coefficient, $D_a$ is the aqueous-phase molecular diffusion coefficient of O3, H is the effective Henry’s law constant of O$_3$ (Sander, 2015), R is the ideal gas constant, f$_r$ is the reacto-diffusive correction term (Shao et al., 2019), and K is the pseudo-first order reaction rate constant between S(IV) and O$_3$ in the aqueous phase (Seinfeld and Pandis, 2006).
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Author contributions: V.A.K. and J.L. planned the research, V.A.K., A.P.T. and A.P. performed the model calculations, V.A.K., A.P., and J.L. analyzed the results, V.A.K. and J.L. wrote the paper. All authors contributed to the manuscript.

Competing interests: Authors declare no competing interests. Code/Data availability: Data and related material can be obtained from V.A.K. (v.karydis@fz-juelich.de) upon request.
Figure 1: Mean, near-surface fine aerosol pH during the period 2010-2015 (central panel). Surrounding panels show the temporal pH evolution during the period 1970-2020 at locations defined in Table 1. Black lines represent the reference simulation. Red and blue lines show the sensitivity simulations in which crustal particle and NH\textsubscript{3} emissions are removed, respectively. Ranges represent the 1σ standard deviation. The anomaly in 1991/2 is related to the Mt Pinatubo eruption.
Figure 2: Average seasonal cycle of modelled pH during the period 2010-2015 at locations defined in Table 1. Ranges represent the 1σ standard deviation.
Figure 3: Time evolution of particle phase fraction of total nitrate as a function of pH over Europe (left), the Eastern USA (right) and East Asia (bottom) during the period 1970-2020.
Figure 4: Time evolution of annual average aerosol hygroscopicity (Kappa) as a function of pH over Europe (left), the Eastern USA (right) and East Asia (bottom) during the period 1970-2020 at the lowest cloud-forming level (940 hPa).
Figure 5: Time evolution of the sulfate production rate on aqueous aerosols from the \( \text{SO}_2 + \text{O}_3 \) multiphase chemistry reaction as a function of aerosol pH over East Asia (left) and South Asia (right) during the period 1970-2020.
Table 1: Decadal averages of aerosol pH.

| Region           | Longitude | Latitude | 1971-1980 | 1981-1990 | 1991-2000 | 2001-2010 | 2011-2020 |
|------------------|-----------|----------|-----------|-----------|-----------|-----------|-----------|
| Western USA¹     | 90°-70°W  | 30°-46°N | 4.6       | 4.8       | 4.8       | 5.0       | 5.1       |
| Eastern USA¹     | 124°-114°W | 30°-52°N | 2.2       | 2.4       | 2.4       | 2.9       | 3.3       |
| Central America¹ | 106°-52°W | 4°-28°N  | 4.6       | 4.6       | 4.6       | 4.7       | 4.9       |
| Europe¹          | 12°W-36°E | 34°-62°N | 2.8       | 3.0       | 3.3       | 3.7       | 3.9       |
| East Asia¹       | 100°-114°E | 20°-44°N | 5.3       | 5.2       | 5.1       | 4.7       | 4.5       |
| South Asia¹      | 68°-94°E  | 8°-32°N  | 5.6       | 5.5       | 5.3       | 5.0       | 4.9       |
| South America¹   | 75°-35°W  | 30°-0°S  | 5.2       | 5.1       | 5.1       | 5.1       | 5.1       |
| Central Africa¹  | 10°-40°E  | 10°S-10°N | 4.9       | 4.8       | 4.8       | 4.7       | 4.9       |
| Southeast Asia¹  | 94°-130°E | 12°S-20°N | 4.5       | 4.3       | 4.1       | 3.9       | 3.8       |
| Middle East¹     | 36°-60°E  | 12°-34°N | 7.0       | 7.0       | 6.9       | 6.9       | 6.8       |
| Arctic           | 0°-360°   | 60°-90°N | 4.2       | 4.2       | 4.6       | 4.8       | 5.2       |
| North extratropics² | 0°-360° | 20°-60°N | 4.8       | 4.8       | 4.7       | 4.7       | 4.9       |
| Tropical oceans² | 0°-360°   | 20°S-20°N | 5.6       | 5.6       | 5.5       | 5.5       | 5.5       |
| South extratropics² | 0°-360° | 60°-20°S | 6.8       | 6.8       | 6.8       | 6.8       | 6.8       |
| Antarctic        | 0°-360°   | 90°-60°S | 5.9       | 5.9       | 5.6       | 5.8       | 5.8       |

¹Only values over land are considered for the calculation of pH
²Only values over oceans are considered for the calculation of pH
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Figure S1: Time evolution of annual average pH as a function of cation/anion molar ratio over Europe (left) and the Eastern USA (right) during the period 1970-2020.

Figure S2: Temporal pH evolution in East and South Asia during the period 2008-2020. Black lines represent the reference simulation. Red lines show the sensitivity simulation in which SO$_2$ emissions are reduced by 75% in East Asia and increased by 50% in South Asia. Ranges represent the 1σ standard deviation.
Figure S3: Absolute change in the calculated mean near-surface fine aerosol pH during the period 2010-2015 (cf. central panel in Fig. 1) by assuming that aerosols are always aqueous solution droplets (metastable state). A negative change corresponds to more acidic particles compared to the stable state assumption.
Table S1: Fractional emission factors of aerosol components for biofuel combustion, and savannah and tropical forest biomass burning (Akagi et al., 2011), and for sea salt (Seinfeld and Pandis, 2006).

| Location     | Latitude | Longitude | Time period         | Simulated mean pH | Field derived mean pH | Method used                  | Reference          |
|--------------|----------|-----------|---------------------|-------------------|-----------------------|------------------------------|--------------------|
| Pellston, MI, USA | 45.55°N | 84.78°W   | Jul 2016            | 3.8               | 3.5                   | pH indicator-paper/colorimetric-image | Craig et al., 2018 |
| Ann Arbor, MI, USA | 42.28°N | 83.74°W   | Aug 2016            | 4.3               | 3.5                   | pH indicator-paper/colorimetric-image | Craig et al., 2018 |
| Centreville, AL, USA | 32.9°N  | 87.25°W   | Jun 1998—Aug 2013   | 6.4               | 1.2                   | ISORROPIA (no NH₃)          | Weber et al., 2016 |
| Centreville, AL, USA | 32.9°N  | 87.25°W   | Jun—Jul 2013        | 7.0               | 1.1                   | ISORROPIA                  | Pye et al., 2018   |
| Egbert, ON, Canada | 44.23°N | 79.78°W   | Jul—Sep 2012        | 3.9               | 2.1                   | E-AIM Model II             | Murphy et al., 2012 |
| Harrow, ON, Canada | 42.03°N | 82.30°W   | Jun—Jul 2007        | 4.2               | 1.6                   | E-AIM Model II             | Murphy et al., 2017 |
| Pendeleen, CA, USA | 34.14°N | 118.42°W  | Jun 2010            | 5.0               | 3.7                   | ISORROPIA (metastable)     | Guo et al., 2017   |
| Toronto, Canada   | 43.66°N | 79.40°W   | 2007-2010           | 4.6               | 2.6                   | E-AIM I (with gas NH₃ HNO₃) | Tao and Murphy, 2019 |
| Toronto, Canada   | 43.66°N | 79.40°W   | 2014-2016           | 4.4               | 2.7                   | E-AIM I (with gas NH₃ HNO₃) | Tao and Murphy, 2019 |
| Ottawa, Canada    | 45.43°N | 75.68°W   | 2007-2016           | 4.0               | 2.5                   | E-AIM I (with gas NH₃ HNO₃) | Tao and Murphy, 2019 |
| Simcoe, Canada    | 44.86°N | 80.27°W   | 2007-2016           | 4.4               | 2.41                  | E-AIM I (with gas NH₃ HNO₃) | Tao and Murphy, 2019 |
| Montreal, Canada  | 45.65°N | 73.57°W   | 2007-2016           | 4.0               | 2.4                   | E-AIM I (with gas NH₃ HNO₃) | Tao and Murphy, 2019 |
| Windsor, Canada   | 42.29°N | 83.07°W   | 2007-2010           | 4.4               | 2.1                   | E-AIM I (with gas NH₃ HNO₃) | Tao and Murphy, 2019 |
| Windsor, Canada   | 42.29°N | 83.07°W   | 2012-2016           | 4.5               | 2.4                   | E-AIM I (with gas NH₃ HNO₃) | Tao and Murphy, 2019 |
| St. Anicet, Canada | 45.12°N | 74.20°W   | 2007-2016           | 4.0               | 2.5                   | E-AIM I (with gas NH₃ HNO₃) | Tao and Murphy, 2019 |
| Sao Paulo, Brazil  | 23.55°S | 46.63°W   | Aug—Sep 2012        | 6.2               | 4.8                   | E-AIM                     | Vieira-Filho et al., 2016 |
| Location          | Latitude | Longitude | Period               | EF | NOx | Model               | Authors          |
|-------------------|----------|-----------|----------------------|----|-----|---------------------|------------------|
| Po Valley, Italy  | 45.40°N  | 12.20°E   | Mar 2009—Jan 2010    | 4.5| 3.1 | E-AIM Model IV      | Squizzato et al., 2013 |
| Po Valley, Italy  | 45.40°N  | 12.20°E   | Spring 2009          | 4.3| 3.6 | E-AIM Model IV      | Squizzato et al., 2013 |
| Po Valley, Italy  | 45.40°N  | 12.20°E   | Summer 2009          | 4.8| 2.3 | E-AIM Model IV      | Squizzato et al., 2013 |
| Po Valley, Italy  | 45.40°N  | 12.20°E   | Fall 2009            | 4.5| 3   | E-AIM Model IV      | Squizzato et al., 2013 |
| Po Valley, Italy  | 45.40°N  | 12.20°E   | Winter 2009-2010     | 4.4| 3.4 | E-AIM Model IV      | Squizzato et al., 2013 |
| Po Valley, Italy  | 45.40°N  | 12.20°E   | Winter 2012-2013     | 4.2| 3.9 | ISORROPIA           | Masiol et al., 2013 |
| Po Valley, Italy  | 45.40°N  | 12.20°E   | Spring 2012          | 4.1| 2.3 | ISORROPIA           | Masiol et al., 2013 |
| Cabauw, Netherlands| 51.97°N  | 4.93°E    | Jul 2012—Jun 2013    | 4.0| 3.7 | ISORROPIA           | Guo et al., 2018   |
| Cabauw, Netherlands| 51.97°N  | 4.93°E    | Jun—Aug 2013         | 3.6| 3.3 | ISORROPIA           | Guo et al., 2018   |
| Cabauw, Netherlands| 51.97°N  | 4.93°E    | Dec—Feb 2012         | 4.1| 3.9 | ISORROPIA           | Guo et al., 2018   |
| Beijing, China    | 39.99°N  | 116.30°E  | Nov 2015—Dec 2016    | 4.8| 4.3 | ISORROPIA           | Liu et al., 2017   |
| Guangzhou, China  | 23.13°N  | 113.26°E  | Jul 2013             | 2.6| 2.5 | E-AIM Model IV      | Jia et al., 2018   |
| Beijing, China    | 39.97°N  | 116.37°E  | Nov 2014—Dec 2014    | 4.5| 4.6 | ISORROPIA           | Song et al., 2018  |
| Beijing, China    | 40.41°N  | 116.68°E  | Oct 2014—Jan 2015    | 5.6| 4.7 | ISORROPIA           | He et al., 2018    |
| Beijing, China    | 39.99°N  | 116.31°E  | Jan—Dec 2014         | 4.8| 3.0 | ISORROPIA           | Tan et al., 2018   |
| Beijing, China    | 39.99°N  | 116.31°E  | Winter 2014          | 5.5| 4.1 | ISORROPIA           | Tan et al., 2018   |
| Beijing, China    | 39.99°N  | 116.31°E  | Fall 2014            | 6.0| 3.1 | ISORROPIA           | Tan et al., 2018   |
| Beijing, China    | 39.99°N  | 116.31°E  | Spring 2014          | 5.4| 2.1 | ISORROPIA           | Tan et al., 2018   |
| City, Country | Latitude | Longitude | Season | ISORROPIA Version | Reference |
|--------------|----------|-----------|--------|-------------------|-----------|
| Beijing, China | 39.99°N | 116.31°E | Summer 2014 | 3.1 | ISORROPIA (metastable) | Tan et al., 2018 |
| Tianjin, China | 39.44°N | 117.16°E | Dec 2014—Jun 2015 | 4.4 | ISORROPIA (metastable) | Shi et al., 2017 |
| Tianjin, China | 39.44°N | 117.16°E | Aug 2015 | 4.4 | ISORROPIA (metastable) | Shi et al., 2017 |
| Beijing, China | 39.98°N | 116.28°E | Feb 2017 | 4.7 | ISORROPIA | Ding et al., 2019 |
| Beijing, China | 39.98°N | 116.28°E | Apr—May 2016 | 5.2 | ISORROPIA | Ding et al., 2019 |
| Beijing, China | 39.98°N | 116.28°E | Jul—Aug 2015 | 4.5 | ISORROPIA | Ding et al., 2019 |
| Guangzhou, China | 23.13°N | 113.26°E | Jul—Sep 2013 | 2.7 | E-AIM Model III | Jia et al., 2018 |
| Hohhot, China | 40.48°N | 111.41°E | Summer 2014 | 5.5 | ISORROPIA (metastable, no NH₃) | Wang et al., 2019 |
| Hohhot, China | 40.48°N | 111.41°E | Autumn 2014 | 6.8 | ISORROPIA (metastable, no NH₃) | Wang et al., 2019 |
| Hohhot, China | 40.48°N | 111.41°E | Winter 2014 | 5.8 | ISORROPIA (metastable, no NH₃) | Wang et al., 2019 |
| Hohhot, China | 40.48°N | 111.41°E | Spring 2015 | 6.4 | ISORROPIA (metastable, no NH₃) | Wang et al., 2019 |
| Hohhot, China | 40.48°N | 111.41°E | 2014—2015 | 6.3 | ISORROPIA (metastable, no NH₃) | Wang et al., 2019 |
| Beijing, China | 40.41°N | 116.68°E | Oct 2014—Jan 2015 | 5.6 | ISORROPIA (stable state) | He et al., 2018 |
| Xi’an, China | 34.23°N | 108.89°E | Nov—Dec 2012 | 5.7 | ISORROPIA | Wang et al., 2016 |
| Beijing, China | 39.99°N | 116.30°E | Jan—Feb 2015 | 5.0 | ISORROPIA | Wang et al., 2016 |
| Beijing, China | 40.35°N | 116.30°E | Jun—Aug 2005 | 4.2 | E-AIM Model II (only aerosols) | Pathak et al., 2009 |
| Shanghai, China | 31.45°N | 121.10°E | May—Jun 2005 | 3.5 | E-AIM Model II (only aerosols) | Pathak et al., 2009 |
| Location          | Latitude | Longitude | Date Range          | E-AM Model | E-AM Model Details | Citation        |
|-------------------|----------|-----------|---------------------|------------|--------------------|----------------|
| Lanzhou, China    | 36.13°N  | 103.68°E  | Jun–Jul 2006        | 6.8        | 0.6                | Pathak et al., 2009 |
| Beijing, China    | 40.32°N  | 116.32°E  | Jan 2005–Apr 2006   | 5.1        | 0.7                | He et al., 2012   |
| Chongqing, China  | 29.57°N  | 106.53°E  | Jan 2005–Apr 2006   | 3.6        | 1.5                | He et al., 2012   |
| Beijing, China    | 40°N     | 116.33°E  | Jan 2013            | 4.6        | 5.8                | Wang et al., 2016 |
| Singapore         | 1.3°N    | 103.78°E  | Sep–Nov 2011        | 3.2        | 0.6                | Behera et al., 2013 |
| Hong Kong         | 22.34°N  | 114.26°E  | Jul 1997–May 1998   | 3.3        | 0.3                | Yao et al., 2002  |
| Hong Kong         | 22.34°N  | 114.26°E  | Nov 1996–Nov 1997   | 3.4        | 1.1                | Yao et al., 2002  |
| Hong Kong         | 22.34°N  | 114.26°E  | Oct 2008            | 5.0        | 0.6                | Xue et al., 2011  |
| Hong Kong         | 22.34°N  | 114.26°E  | Nov 2008            | 3.2        | 0.5                | Xue et al., 2011  |
| Hong Kong         | 22.34°N  | 114.26°E  | Jun–Jul 2009        | 1.6        | 0.1                | Xue et al., 2011  |
| Pacific Ocean     | 47.5°S   | 147.5°W   | Nov–Dec 1995        | 7.0        | 1.0                | Fridlind and Jacobson, 2000 |
| South Ocean       | 61°S     | 45°W      | Jan 2015            | 6.9        | 1.4                | Dall’Osto et al., 2019 |
| South Ocean       | 64°S     | 65°W      | Jan–Feb 2015        | 6.9        | 3.8                | Dall’Osto et al., 2019 |
Table S2: Fractional emission factors of aerosol components for biofuel combustion, and savannah and tropical forest biomass burning (Akagi et al., 2011), and for sea salt (Seinfeld and Pandis, 2006).

| Source                  | SO$_4^{2-}$ | NO$_3^-$ | Cl$^-$ | Na$^+$ | K$^+$ | Mg$^{2+}$ | Ca$^{2+}$ | NH$_4^+$ |
|-------------------------|-------------|----------|--------|--------|-------|-----------|-----------|---------|
| Biofuel combustion      | -           | 0.014    | -      | 0.003  | 0.023 | 0.073     | -         | -       |
| Grassfire burning       | 0.05        | 0.04     | 0.62   | 0.01   | 0.02  | 0.04      | 0.06      | 0.01    |
| Forest fire burning     | 0.25        | 0.21     | 0.29   | 0.04   | 0.56  | 0.08      | 0.16      | 0.01    |
| Sea salt                | 0.077       | -        | 0.55   | 0.306  | 0.011 | 0.037     | 0.012     | -       |

Table S3: Fractional chemical composition of mineral dust emissions (Karydis et al., 2016).

| Desert                  | Na$^+$     | K$^+$     | Ca$^{2+}$ | Mg$^{2+}$ | Other |
|-------------------------|------------|-----------|-----------|-----------|-------|
| Great Basin             | 0.064      | 0.023     | 0.053     | 0.018     | 0.842 |
| Mojave                  | 0.015      | 0.027     | 0.050     | 0.019     | 0.880 |
| Sonoran                 | 0.025      | 0.042     | 0.037     | 0.006     | 0.020 |
| Patagonia               | 0.012      | 0.035     | 0.021     | 0.013     | 0.020 |
| Monte                   | 0.023      | 0.018     | 0.025     | 0.009     | 0.025 |
| Atacama                 | 0.069      | 0.007     | 0.018     | 0.005     | 0.901 |
| Kalahari                | 0.030      | 0.050     | 0.120     | 0.090     | 0.710 |
| Namibia                 | 0.011      | 0.035     | 0.075     | 0.030     | 0.849 |
| Sahara                  | 0.014      | 0.004     | 0.014     | 0.006     | 0.946 |
| Saudi Arabia            | 0.022      | 0.033     | 0.082     | 0.022     | 0.841 |
| Thar/Lut                | 0.023      | 0.040     | 0.120     | 0.028     | 0.840 |
| Taklimakan              | 0.012      | 0.021     | 0.077     | 0.017     | 0.873 |
| Gobi                    | 0.028      | 0.004     | 0.005     | 0.003     | 0.964 |
| Great Sandy/Simpson     | 0.012      | 0.015     | 0.024     | 0.000     | 0.940 |
| Other                   |            |           |           |           |       |