Estimation of nutrient contributions from the ocean across a river basin using stable isotope analysis

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Abstract

Total nitrogen (TN), which consists of total particulate nitrogen (TPN) and total dissolved nitrogen (TDN), is transported with not only in river channels but also across the entire river basin, including via ground water and migratory animals. In general, TPN export from an entire river basin to the ocean is larger than TDN in a mountainous region. Since marine derived nutrients (MDN) are hypothesized to be mainly transported as suspended matters from the ground surface, it is necessary to investigate the contribution of MDN to the forest floor (soils) in order to quantify the true role of MDN at the river ecosystem scale. This study investigated TN export from an entire river basin, and also we estimated the contribution of pink (*Oncorhynchus gorbuscha*) and chum salmon (*O. keta*) to total oceanic nitrogen input across a river basin. The potential contribution of TN entering the river basin by salmon was found to be 23.8 % relative to the total amount of TN exported from the river basin. The contribution of particulate nitrogen based on suspended sediment from the ocean to the river basin soils was 22.9 % with SD of 3.6 % by using stable isotope analysis (SIA) of nitrogen ($\delta^{15}$N). Furthermore, SIA showed that the transport of oceanic TN by sea eagles (*Haliaeetus* spp.) was greater than that by bears (*Ursus arctos*), which had previously been thought bears are thought to be the major animal transporter of nutrients in the northern part of Japan.
1. Introduction

SIA is increasingly being used to examine connectivity in coastal aquatic-terrestrial ecosystems, such as the input of MDN from the open ocean to coastal and widely river ecosystems (Wyatt et al., 2010a; Wyatt et al., 2010b; Wyatt et al., 2012). In the case of river ecosystems, the transportation of nutrients, such as nitrogen and phosphorus, by migrating fish results in enhancement of biofilms and planktonic productivity in river systems (Juday et al., 1932; Cederholm and Peterson, 1985; Bilby et al., 1996; Gresh et al., 2000; Chaloner et al., 2002; Moore and Schindler, 2004; Yanai and Kochi, 2005; Levi and Tank, 2013). Most of those cases, many terrestrial consumers like mammals, birds, fishes and insects have been shown to play a large role in terms of providing MDN to watersheds (Donaldson, 1966; Ben-David et al., 1997a; Hilderbrand et al., 1999; Gende et al., 2002; Naiman et al., 2002; Wilkinson et al., 2005; Bartz and Naiman, 2005). Moreover, MDN inputs have been shown as important processes controlling the productivity of ecosystem. For example, Merz and Moyle (2006) found that the contribution of MDN to the foliar nitrogen of wine grapes was about 18 to 25 %. Also, Hilderbrand et al. (1999) demonstrated that trees and shrubs near spawning streams receive 24 to 26 % of the foliar nitrogen from MDN, while Helfield and Naiman (2002) suggested that 15.5 to 17.8 % of spruce foliage nitrogen is provided from MDN. Thus, isotopic methods as intrinsic geospatial tracer provided quantification of
cross-ecosystem transfer of nutrients. In particular, migrating fish, such as salmon, have been found to be necessary for a sustainable nutrient-cycle system due to their important role as nutrient transporters (Ben-David et al., 1998; Wipfli et al., 1998; Yanai and Kochi, 2005; Gende et al., 2007; Hocking and Reimchen, 2009; Hocking and Reynolds, 2011). Additionally, MDN has been demonstrated to be important not only for river ecosystems but also potentially for upstream lakes (Kline et al., 1990; Kline et al., 1993; Schindler et al., 2003).

Here we focus on the particulate nutrient budgets due to suspended sediment at the river basin scale using stable isotopes. When we consider nutrient flux in a river flowing from the upstream end into the ocean, the flux depends on nutrients supplied not only inside the river itself but also from the entire river basin (Dutta and Nakayama, 2010; Alam and Dutta, 2012; Riggsbee et al., 2008). Also, particulate nutrient flux, which is given from surface soils dominantly, is revealed to be larger than dissolved nutrient generally in a mountainous region (Nakayama et al., 2011). Cederholm et al. (1989) demonstrated that mammals and birds consume migrating fish, which may result in the secondary dispersion of MDN across the river basin associated with the movement of these consumers. Other studies have revealed that mammals incorporate MDN from salmon, which may subsequently lead to re-export to the ocean through river flows
(Bilby et al., 1996; Ben-David et al., 1997a; Ben-David et al., 1997b; Hilderbrand et al., 1999; Szepanski et al., 1999; Reimchen, 2000). However, the contribution of MDN to surface soils, which may be transported from a river basin to the ocean as suspended sediments, at the river basin scale has not been adequately quantified in natural systems because of difficulty to show those complex food web and accurate biomass.

In this study we present the TN transport across an entire river basin to the ocean, the potential contribution of TN from the ocean to a river basin by salmon, and the contribution of MDN to surface soils in a river basin. Integrated stable isotope researches in the geological, hydrological and biological aspects allowed us to estimate nutrient budgets in natural river basin and convinced us to conserve the ocean river connectivity.

2. Geophysical setting

Our target area, the Shiretoko Peninsula, was registered as a World Natural Heritage area in July of 2005. Shiretoko is located at the southernmost extent of drift ice and its ecological systems exhibit high biodiversity and high rates of nutrient circulation, particularly due to runs of pink (Oncorhynchus gorbuscha) and chum (O. keta) salmon from the Sea of Okhotsk. Potential runs of salmon along the coast of Hokkaido in the
Sea of Okhotsk have been estimated at about 29,900,000 individuals a year (Hokkaido National Fisheries Research Institute, Fisheries Research Agency, 2009), equivalent to 2,590 tons of total nitrogen. The size of the Okhotsk coastal region of Hokkaido is about 24,000 km², which corresponds to that the mean total nitrogen input from the ocean is about 108 kg km⁻² yr⁻¹ if we assume that all salmon run up rivers and the total nitrogen is distributed into the river basins completely. Shiretoko is located on the northeast coast of Hokkaido, Japan (approximately 43°57' N to 44°21' N and 144°58' E to 145°23'E), and has a width, length and maximum altitude of about 15 km, 50 km and 1660 m, respectively (Fig. 1). The Rausu River basin was selected as a study area because its watershed is the largest in the region and it is considered a representative watershed in the Shiretoko Peninsula. The watershed area, river length, and the mean river slope are 32.5 km², 7 km, 1/7, respectively (Fig. 2). Because of the steep slope, nutrient flux due to suspended sediments is larger than dissolved nutrient flux in the Rausu River basin (Nakayama et al. 2011). Field experiments were carried out over 5 years from 2008 to 2012.

3. Methods

3.1 Nitrogen from a river basin to the ocean
TN, TDN and TPN were measured at St.0 around the river mouth from 2007 to 2009 (Fig. 2). The nitrogen concentration of filtered and non-filtered water samples were analyzed by the cadmium reduction-colorimetric method. Annual TN and annual TDN exports to the ocean were evaluated using the river discharge at St.0 with TDN-discharge and TPN-discharge curves. The TDN-discharge and TPN-discharge curves were produced using ten different peak discharge floods and base flow discharges. As river discharge was not measured during the winter season from January to March, a storage function method was applied to estimate river discharge from 2008 to 2012 (Michael, 1978; Michael et al., 1979). The validity of the storage function method was confirmed through comparison with the observed river discharge from April to December.

3.2 Salmon runs

To evaluate the contribution of salmon to soil organic matter (SOM), salmon runs were investigated in the Rausu River. Salmon were caught at the river mouth for artificial incubation and release, providing an estimate of the number of salmon caught by the apparatus (Hokkaido National Fisheries Research Institute, Fisheries Research Agency, 2009). The apparatus for catching salmon consisted of lattice fence, which does not obstruct flood flow or completely block the runs of salmon. Therefore, it was necessary
to quantify the capture rate of the apparatus in order to estimate the actual volume of salmon runs. Field observations were conducted in the Tokorohoronai River, which is located in the same region of Hokkaido and has a custom to remove its apparatus before and after salmon run seasons, allowing us to monitor the salmon escapement from the apparatus and the salmon run under the open condition at the same place. The capture rate of the apparatus was calculated with numbers of salmons passing the observation point which has a channel section of 3 m in width and 0.2 m in depth, instead of the Rausu River because its river width (about 15 m) is too wide to cover the entire width. We used two infrared cameras (SM-AVIR-602S, Hero Corp., Izumo, Japan) placed 2 m above the river surface and recorded videos in all day to monitor the individual salmon passing this 3 m section. Videos were taken from the 25th to 28th of November (before removal of the apparatus) and from the 4th to 7th of December (after removal of the apparatus) in 2013. There was no influence of rainfall during the observation period.

3.3 Stable isotope analysis

MDN, such as nitrogen, are generally supplied from the ocean to surface soils in a river basin as SOM, which includes feces of mammals, droppings of birds, and the remains of salmon preyed upon by mammals, birds and insects. To focus on the influence of SOM on particulate nitrogen in the river basin soils, soil particles with diameter of less than
500 µm after rinsing in 1N-HCL solution were used in the analysis. In general, some proportion of the nitrogen is reduced into the atmosphere due to denitrification, which indicates the difficulty evaluating total amount of supplied nitrogen. In this study, it cannot be allowed to evaluate how much dissolved nitrogen is decomposed from SOM. However, TPN export from an entire river basin is revealed to be larger than TDN in the Rausu River basin due to the steep slope (Nakayama et al., 2011). Therefore, we made an attempt to estimate the contribution of MDN to SOM as a sequel to an accumulation, which directly corresponds to the suspended sediments transporting particulate nutrient through a river to the ocean, by sampling surface soils across the Rausu River basin (Fig. 2). Surface soil samples were taken at 12, 20 and 21 stations in 2008, 2009 and 2012, respectively. In 2008, fewer samples were taken as we did not have permission to sample surface soils in special protection zones. Surface soils were sampled from three different points at each station in a volume of 15 cm × 15 cm × 5 cm (height × width × depth). Surface soil sampling stations in 2012 are shown in Fig. 2. Since previous studies have revealed that surface soil transport is related to the spatial distribution of surface soil type, land-use type and vegetation (Ishida et al., 2010), the location of each sampling station was selected by dividing the river basin into 21 domains (sub-basin areas) that vary in soil type and vegetation. The spatial distribution of surface soil type is divided into 6 categories. Although the spatial pattern in vegetation is complicated,
the vegetation can generally be categorized in terms of altitude. Since Shiretoko is protected as natural World Heritage area, all areas studied are classified as forest and have high vegetation cover. Stable isotope ratios of carbon ($\delta^{13}$C) and nitrogen ($\delta^{15}$N) were measured using a Delta Plus Advantage mass spectrometer (Thermo Electron) coupled with an elemental analyzer (Flash EA 1112, Thermo Electron) at the Port and Airport Research Institute, Japan. Stable isotope ratios are expressed in $\delta$ notation as the deviation from standards in parts per thousand (‰) according to the following equation:

$$\delta^{13}$C, $\delta^{15}$N = \left[\frac{R_{\text{sample}}}{R_{\text{standard}}} - 1\right] \text{ (1)}$$

where $R = \frac{^{13}$C}{^{12}$C}$ or $\frac{^{15}$N}{^{14}$N}$. PeeDee Belemnite and atmospheric nitrogen were used as the isotope standards of carbon and nitrogen, respectively. The analytical precision in the mass spectrometer system based on the standard deviation of the internal reference (L-histidine) replicates was <0.15‰ for both $\delta^{13}$C and $\delta^{15}$N. The contribution of MDN to SOM in surface soils was evaluated by applying a two source mixing model based on stable isotope analysis (SIA) of carbon ($\delta^{13}$C) and nitrogen ($\delta^{15}$N) (Kline et al., 1998; Moore and Semmens, 2008; Hossler and Bauer, 2012). Three soil samples were collected at each sampling.
station in order to account for small scale variability in SOM (Fig. 2). Salmon tissue isotopes were considered representative of the isotope composition of ocean productivity. To isotopically characterize terrestrial productivity, we considered one terrestrial end-members (sources): Soil Samples exhibiting the Lowest values of $\delta^{13}$C and $\delta^{15}$N (hereafter SSL), and thus assumed to have the highest terrestrial contribution to SOM. SSL was collected close to the top of the mountain, where MDN is not expected to influence isotope values. Representative soil samples collected in the same river basin were chosen because they have isotopically similar characteristics to the target soil samples in this study.

The contribution of MDN to SOM was evaluated using a two sources mixing model based on the measured $\delta^{13}$C and $\delta^{15}$N. The average contribution in the Rausu River basin was computed using each sub-basin area obtained from the Thiessen method.

\[
\begin{align*}
    f_{\text{C-MDN}} + f_{\text{C-LDN}} &= 1 \\
    f_{\text{C-MDN}} \delta^{13}\text{C}_{\text{salmon}} + f_{\text{C-LDN}} \delta^{13}\text{C}_{\text{SSL}} &= \delta^{13}\text{C}_{\text{soil}} \\
    f_{\text{N-MDN}} + f_{\text{N-LDN}} &= 1 \\
    f_{\text{N-MDN}} \delta^{15}\text{N}_{\text{salmon}} + f_{\text{N-LDN}} \delta^{15}\text{N}_{\text{SSL}} &= \delta^{15}\text{N}_{\text{soil}}
\end{align*}
\]
where $f_{C, MDN}$ and $f_{C, LDN}$ are the contributions of MDN and land-derived nutrient (LDN) by carbon, $\delta^{13}C_{\text{salmon}}$, $\delta^{13}C_{\text{SSL}}$ and $\delta^{13}C_{\text{soil}}$ are the stable isotope ratios of carbon for salmon, SSL and soil samples, respectively, $f_{N, MDN}$ and $f_{N, LDN}$ are the contributions of MDN and LDN by nitrogen, $\delta^{15}N_{\text{salmon}}$, $\delta^{15}N_{\text{SSL}}$ and $\delta^{15}N_{\text{soil}}$ are the stable isotope ratios of nitrogen for salmon, SSL and soil samples, respectively.

As bamboo grass (*Sasa senanensis*) is the dominant species in the study area, bamboo grass was collected at 6 soil sampling points. Furthermore, to investigate the contribution of the other typical mammals and birds in Shiretoko to SOM, and to roughly understand these animals diets, droppings of sea eagles (*Haliaeetus* spp.) and feces of brown bear (*Ursus arctos*) were collected (Kuwae et al., 2008; Kuwae et al., 2012). Chum salmon tissues and droppings of sea eagles were collected at the river mouth and feces of brown bear were collected at St.14. Then those samples were used into SIA after rinsing in chloroform-methanol solution (2:1). Dropping and feces samples can reduce the uncertainty in terms of SIA fractionation factors when compared to tissue samples (e.g., muscle, liver, and blood). As fractionation occurs during the making or breaking of bonds in small molecules, we might not expect fractionation during food assimilation, i.e., uptake of large molecules, in the absence of the breaking of nitrogen bonds (Fry, 2006). Thus, while tissue samples have variability and
uncertainty related to fractionation factors (body conditions such as fasting), we consider that feces and dropping samples do not. However, in the case of multiple food sources, feces and dropping are likely to be enriched in relatively indigestible food sources, when compared with stomach contents or assimilated materials (Sponheimer et al., 2003; Kuwae et al., 2008). A further advantage of using droppings, as opposed to tissues, is that no killing and/or damage to wildlife is involved in collecting samples. Multivariate analysis of variance (MANOVA) and post hoc tests by Tukey-Kramer were used to investigate differences in surface soil samples.

4. Results and Discussion

4.1 Estimation of nitrogen export to the ocean

During 2007 to 2009 the concentration of TDN was observed to be constant, 0.090 mg L\(^{-1}\) (SD 0.022 mg L\(^{-1}\)), regardless of the discharge in the Rausu River. In contrast, TPN was revealed to be a function of river discharge (\(r^2=0.88\); Eq. 6) (Fig. 3). TPN showed a strong correlation with suspended sediment (SS) concentrations, with SS concentration increasing with increasing river discharge (Fig. 3). TPN was modeled by using our field observation results, discharge and TPN as (6).

\[
TPN = 0.0032 \times Q^{1.771}
\]  

(6)
where $Q$ is the river discharge ($m^3 \cdot s^{-1}$).

The validity of the storage function method model was confirmed using the observed river discharge from April to September of 2009, which resulted in a Coefficient of Determination (CoD) of 0.61. The reliability of the model has been shown to be high enough for the analysis of river discharge when the CoD is more than 0.6 (Dutta and Nakayama, 2010). Annual mean export of TDN, TPN and TN from 2008 to 2012 were 5210 kg yr$^{-1}$, 14750 kg yr$^{-1}$ and 19960 kg yr$^{-1}$, respectively. Since the size of the Rausu River basin of Shiretoko is 32.5 km$^2$, the annual mean exports of TDN, TPN and TN per unit catchment area were 160 kg km$^{-2}$ yr$^{-1}$, 454 kg km$^{-2}$ yr$^{-1}$ and 614 kg km$^{-2}$ yr$^{-1}$, respectively (Table 1). The average concentrations of TDN and TPN from 2008 to 2012 were 0.090 mg L$^{-1}$ and 0.216 mg L$^{-1}$, which agrees with a previous study at the site (Nakayama et al., 2011).

4.2 Contribution of salmon runs to nitrogen input from the ocean

The average number of salmon passing the cameras in the Tokorohoronai River during the 4 days while the apparatus for catching salmon was present was 0.49 hr$^{-1}$. The average numbers for 4 days after the apparatus was removed from the river was 0.61
hr\(^{-1}\), so the rate of capture of salmon by the apparatus was estimated as 20%:

\[
\frac{(0.61-0.49)}{0.61} = 0.20.
\]

In the Rausu River of Shiretoko, the annual average numbers of salmon caught by the apparatus at the river mouth were 3075 and 10580 for chum and pink salmon, respectively, from 2001 to 2009. By assuming that all apparatuses have the same rate of capture, the potential for chum and pink salmon runs can be estimated as 15375 and 52900, respectively. The average weight of chum and pink salmon are 3.3 kg and 2.0 kg, respectively (Makiguchi et al., 2007), with a nitrogen content of about 30.4 g kg\(^{-1}\) (Larkin and Slaney, 1997). Therefore, annual TN potentially transported by chum and pink salmon is estimated to be 1542 kg yr\(^{-1}\) and 3216 kg yr\(^{-1}\), respectively. Finally, the annual TN transported by chum and pink salmon per unit catchment area can be estimated as 146 kg km\(^{-2}\) yr\(^{-1}\) (SD 19 kg km\(^{-2}\) yr\(^{-1}\)), which corresponds to the contribution of TN by salmon, 23.8 % (SD 3.1 %), relative to the annual outflow of TN per unit area (considered to be 100 %) (Table 1).

4.3 Contribution of MDN to SOM in the Rausu River basin

The 2012 field experiment suggested that stable isotope ratios, $\delta^{13}C$ and $\delta^{15}N$, were relatively higher close to the ocean (stations 20 and 21) compared to the top of the
mountain (station 19), which has been also demonstrated by Kline et al. (1998) (Fig. 2).

Tissue $\delta^{13}$C and $\delta^{15}$N were highest in the sea eagles (n=8) and lowest in the bamboo grass (n=3). The $\delta^{13}$C and $\delta^{15}$N of SOM (n=53) lay between sea eagles and bamboo grass (Fig. 4). MANOVA suggested that there was no isotopic ($\delta^{13}$C and $\delta^{15}$N) difference between salmon tissue (n=12) and sea eagle droppings. Feces from brown bears (n=7), which were previously thought to be the major terrestrial consumer of spawning salmon, were significantly lower than those of salmon and sea eagles (bear vs. salmon ($\delta^{13}$C): $P < 0.001$; bear vs. salmon ($\delta^{15}$N): $P < 0.001$; bear vs. sea eagle ($\delta^{13}$C): $P < 0.001$; bear vs. sea eagle ($\delta^{15}$N): $P < 0.001$). The stable isotope ratios in sea eagle droppings were the highest among the animals measured. Also, the contributions of the other predators may impact re-export of nutrient from the ocean across the river basin; their role in marine-derived nitrogen input and re-export, such as through release of MDN-rich feces, should be investigated in more detail in future studies.

Both $\delta^{13}$C and $\delta^{15}$N of SOM were lower than those of salmons (Fig. 4). Interestingly, SSL has almost the same value of the mean value of bamboo grass, which may suggest that bamboo grass can be considered to be as LDN. The stable isotope ratios in sea eagle droppings and brown bear feces were higher than LDN. Since brown bears are previously thought to be the major terrestrial consumer of spawning salmon, they may
impact re-export of nutrient from the ocean across the river basin, such as through release of MDN-rich urine and feces (Hilderbrand et al. 1999). Although fractionation occurs due to breaking of bonds in small molecules, we might not expect fractionation during food assimilation in the absence of the breaking of nitrogen bonds (Fry, 2006). Therefore, feces and dropping are likely to be enriched in relatively indigestible food sources, when compared with stomach contents or assimilated materials (Sponheimer et al., 2003; Kuwae et al., 2008). However, it is difficult from Fig. 4 to decide that sea eagles eat more oceanic fish than bears because bears release more MDN as urine compared to dropping (Hilderbrand et al. 1999).

The isotopic composition of salmon as representative of oceanic $\delta^{15}$N and $\delta^{13}$C were 10.99 and -20.54, respectively. The $\delta^{15}$N and $\delta^{13}$C of SSL were -3.19 and -29.48, respectively. Therefore, the three-year average estimate of the contribution of MDN to SOM for $\delta^{15}$N depending on the choice of terrestrial isotope values was obtained e.g. 22.9 % (SD 3.6%) by using a two sources mixing model (Fig. 5). As the reference, the three-year average estimate of the contribution of MDN to SOM for $\delta^{13}$C was 17.7 % (SD 1.1%) (Fig. 5). Since annual export of TPN per unit area from the Rausu River basin to the ocean was 454 kg km$^{-2}$ yr$^{-1}$, annual re-export of TPN originally derived from the ocean is estimated to be 104 kg km$^{-2}$ yr$^{-1}$ ($= 454$ kg km$^{-2}$ yr$^{-1}$ * 22.9%) (SD 16
$ \text{kg km}^{-2} \text{ yr}^{-1} = 454 \text{ kg km}^{-2} \text{ yr}^{-1} \times 3.6\%)$ based on the contribution of MDN to SOM (Table 1 and Fig. 5). However, it should be noted that this value for MDN re-export is estimated without contribution of marine derived TDN and thus should be considered the minimum annual MDN re-export from the viewpoint of TN.

5. Conclusions

In recent decades, field experiments and stable isotope analyses have been employed to understand the contribution of runs of salmon to river ecosystems. In river ecosystems, runs of salmon are thought to play a large role in the sustainability of nutrient circulation due to their contribution to mammals that incorporate MDN and disperse it across the entire river basin, with the MDN potentially re-exported to the ocean through river flows. In previous studies, the input of TN from the ocean to river basin ecosystems has been actively investigated, since it can control ecosystems in which salmon run upstream for spawning, but the contribution of TN from the ocean across an entire river basin has not been examined in detail. This is despite the fact that waterfalls and the other obstacles, which inhibit runs of salmon, are known to reduce the transport of MDN upstream. Therefore, this study quantifies the role of salmon in transporting MDN across an entire river basin of the Shiretoko World Natural Heritage area using stable isotope analysis.
Annual TN transport estimated for pink salmon was twice that for chum salmon, which suggests that pink salmon play a greater role in the input of TN across the Rausu River basin. The potential contribution of TN by salmon was 23.8 % (SD 3.1 %), while the contribution of MDN to SOM was 22.9 % (SD 3.6 %). Therefore, the annual potential contribution of salmon to TN may be 146 kg km$^{-2}$ yr$^{-1}$ ( = 23.8 %), which provides valuable support for a strong influence of MDN on the ecological systems across this river basin (Table 1 and Fig. 6).
Author contribution
K. Nakayama designed the field experiments and wrote most of the paper. Also, K. Nakayama performed mixing model analysis. Also, Y. Maruya produced the figures using the GIS technical input and carried out runoff analysis. K. Katsuaki helped the river discharge and nitrogen concentration analysis. M. Komata, and K. Katsuaki measured total nitrogen, dissolved total nitrogen and particulate total nitrogen. K. Matsumoto carried out the field experiments of salmon runs and conducted statistical analysis of stable isotopes. T. Kuwae designed the field experiment regarding stable isotopes and carried out stable isotope measurements. All authors read and commented on drafts of this paper.

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Fig. captions:

Fig. 1. Coastline around the Shiretoko Peninsula and the Rausu River basin.

Fig. 2. (a) Elevation of the Rausu River basin. Green circles indicate surface soil sampling stations in September of 2012. Red circles indicates a field observation station for discharge, TDN (total dissolved nitrogen) and TPN (total particulate nitrogen). (b) δ15N and sampling stations in 2012. (c) δ13C and sampling stations in 2012.

Fig. 3. River discharge, total particulate nitrogen and suspended sediment at the river mouth of Rausu River. (a) River discharge and concentration of total particulate nitrogen. (b) Concentration of suspended sediment and concentration of total particulate nitrogen.

Fig. 4. δ13C and δ15N of bamboo grass (Sasa senamnensis), SSL (Soil Samples exhibiting the Lowest values of δ13C and δ15N), soil samples, bear feces (Ursus arctos), salmon (Oncorhynshus keta), and sea eagles droppings (Haliaeetus spp.). The bars indicate the standard deviation.

Fig. 5. Contribution of MDN (marine derived nitrogen) from the ocean to the Rausu River basin in 2008, 2009 and 2012 using the two sources mixing model. (a) Average contributions of MDN based on SSL (Soil Samples exhibiting the Lowest values of...
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Fig. 6. Annual input of TN (total nitrogen) per unit area from the Rausu River basin to the ocean, and annual TN transported by salmon per unit area, relative to the annual outflow of TPN per unit area (considered to be 100 %).
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Table 1. Summary of annual export and re-export of nitrogen per unit area.
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| N export | N re-export | Salmon run (%) | MDN input (%) |
|----------|-------------|----------------|---------------|
|          | N kg·y⁻¹   | N kg·km⁻²·y⁻¹ | N kg·km⁻²·y⁻¹ |               |
| TDN      | 5210        | 160            | -             | -             |
| TPN      | 14750       | 454            | -             | 104 (22.9)    |
| TN       | 19960       | 614            | 146 (23.8)    | -             |

* = (Salmon run)/(N export)

** = (N export) × (MDN contribution = 22.9)