Research Article

Removal of Uranyl Ion from Wastewater by Magnetic Adsorption Material of Polyaniline Combined with CuFe$_2$O$_4$

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The magnetic adsorption material of polyaniline (PANI) with amino functional group combined with CuFe$_2$O$_4$ (CuFe$_2$O$_4$/PANI nanocomposite) has been described in this work. It has been characterized by TEM, XRD, XPS, BET, FTIR, and VSM, respectively. Significantly, it exhibits extremely high maximum adsorption capacity (322.6 mg/g) for removal of uranyl ions from wastewater at a pH of 4. The adsorption process is consistent with the quasisecond-order kinetic equation, and the isotherm and kinetic data are accurately described by the Langmuir isothermal adsorption model. Furthermore, the magnetic CuFe$_2$O$_4$/PANI displays stable adsorption performance for uranyl ions after five cycles of recovery in acid medium, which indicates it possesses good recovery due to its magnetism and excellent regeneration ability for reusability.

1. Introduction

In recent years, nuclear power is widely considered as a sustainable solution to the global demand for energy [1]. With the rapid expansion of the atomic energy industry as well as nuclear science and technology, the development of new green energy such as nuclear energy not only brings huge social and economic benefits to human society but also arise increasingly serious environmental problems. Nuclear energy as one of the clean energy sources, its series of environmental and safety problems brought about by its sustainable development have been paid close attention to nowadays. Meanwhile, due to the increasing discharge of radioactive nuclides in wastewater and the increasingly complex composition, people are paying more and more attention to the harmful problem it arises for human health and the ecological environment as well. Environmental contaminant releases that contain uranium are among the more serious problems that must be confronted by the decontamination programs. Uranium with natural radioactivity is one of the most basic nuclides in the field of nuclear research and the core resource of nuclear energy development. [2]. As one of the most common pollutants, uranium also has the characteristics of strong radioactivity, high carcinogenicity, and toxicity, which will cause great radiation harm to the human organism and the ecological environment. Under oxidizing conditions, dissolved uranium is predominantly in the U (VI) (uranyl ions) form and is potentially quite mobile in the aquatic environment. Therefore, removal of uranyl ions from aqueous solution is becoming of great importance for not only sustainable development of nuclear energy but also the environment protection.

More and more research is focused on the removal and recycling of uranium in wastewater ([3, 4]; Ozdemir and Usanmaz, [5]). At present, the commonly used treatment methods for industrial wastewater containing uranium and heavy metal ions mainly include chemical precipitation method [6], extraction method [7], biological method [8], membrane filtration method [9], and adsorption method [10–12]. Adsorption method, as a type of environmental protection, belongs to a low harmful treatment to remove wastewater of radionuclides and heavy metal ion (lead, mercury, cadmium, chromium, etc.), due to its economic advantage, species diversity, good stability, high applicability,
simple operation, high treatment efficiency, and better regeneration [13–17].

Screening suitable adsorption materials with high adsorption capacity is becoming extremely important to remove uranium efficiently from wastewater. [18–20]. The commonly used adsorption materials are including graphene, carbon nanotubes, some carbon materials, cellulose, polymer materials, kaolin, clay, mineral, organic-inorganic hybrid adsorption materials, ion-imprinted polymers, and other composite materials, but their limitations are focused on not very high adsorption efficiency, high cost, low selectivity, long preparation period, and difficulty of recycling. [21–23] [24]. Therefore, the development and utilization of new high efficiency of uranium adsorption material are one of the most important issues of concern to be solved at present.

Magnetic composite nanomaterials have been widely used in environmental protection, catalysis, and biomedicines and have broad application prospects. [25–28]. A large number of studies have been carried out on the removal of heavy metals in wastewater by using magnetic nanoparticles as magnetic core in the composite adsorption materials. It has also been proved that it has the characteristics of strong magnetism, good adsorption, easy separation, and collection under the action of an external magnetic field [29–33]. More importantly, the functional groups of magnetic nanocomposite which is proper prepared based on different matrix can be designed according to the adsorption selectivity of the adsorption material for specific radionuclides. The advantages of it combined with other materials can complete and improve the comprehensive performance, which would not be obtained by a single component. Many new combination properties can solve the problem much better than the single magnetic nanomaterial showing low adsorption performance for pollutants [34, 35].

Polyaniline (PANI) is a type of polymer compound with high conductivity, which contains a large number of amino and imine groups [36]. As the exists of the abundant amounts of primary and secondary amino groups, PANI is expected to have a strong affinity with heavy metal ions by redox reaction [37, 38]. Due to its low production cost, simple synthesis method, good chemical, and environmental stability, PANI has been widely applied to the adsorption of heavy metal pollutants from wastewater in recent years [39–41]. It has also been found that the magnetic PANI nanocomposite modified with GO composites would attract materials with exceptional ability in the selective extraction of trace U (VI) from aqueous solutions [42, 43]. It highlights the extensive applicability of these novel materials in the selective extraction and separation of trace uranyl ions from aqueous solution in nuclear fuel extraction processes.

Therefore, magnetic adsorption material of polyaniline (PANI) combined with CuFe2O4 (CuFe2O4/PANI nanocomposite) has been synthesized in this work. In the meanwhile, the amino functional groups with complexation for uranyl ions were introduced, which would improve the adsorption capacity of the novel materials. The new type of magnetic CuFe2O4/PANI nanocomposite for trace uranyl ions adsorption capacity in aqueous solution under different operating conditions was studied in detail. It was characterized by TEM, FTIR, XRD, XPS, and BET, respectively, and batch adsorption experiments were carried out as well to study the effects on the uranium adsorption of various experimental parameters, such as pH value, contact time, temperature, and initial concentration. The adsorption kinetics, isotherm models, and thermodynamics were evaluated in detail to understand the adsorption process. Finally, the recycling regeneration of the adsorption nanocomposite for its reusability was analyzed in detail through the experiments of adsorption/desorption cycle regeneration.

2. Experimental

2.1. Materials and Reagents. The chemical reagents used in this study were all of analytical grade with 98-99% purity. It includes copper sulfate (CuSO4·5H2O), ferric chloride (FeCl3·6H2O), polyethylene glycol 6000 (PEG), potassium hydroxide (KOH), hydrochloric acid (HCl), aniline (C6H5N), ethyl alcohol (C2H5OH), and ammonium thiosulfate ((NH4)2S2O3). All the analytical grade chemical reagents used in this study were without further purification. Freshly deionized water was used as the dispersion medium. The stock uranium (U (VI)) solution was prepared by dissolving the appropriate amounts of uranyl nitrate hexahydrate [UO2(NO3)2·6H2O] in deionized water. The concentration of uranium stock solution is 1000 mg/L, which is diluted to 5 ~ 300 mg/L by gradient dilution method.

2.2. Instrumentation. Surface morphological images were characterized by transmission electron micrograph (TEM) (FEI Tecnai G2) at an accelerating voltage of 200 kV. The functional groups of the materials were confirmed by Fourier transform infrared (FT-IR) (Vertex70, Bruker company). The materials were pelleted using KBr prior to FTIR measurement. XRD analysis was carried out on a D8 Advance instrument (Bruker company) with Cu Kα radiation source (λ = 1.5406 Å). The operating condition was at a scan rate of 3/min from 10° to 90° to obtain the XRD pattern. Surface areas of samples were determined from N2 adsorption isotherms at 77.3 K with the BET method using a Micromeritics ASAP 2010 analyzer prior to the surface area measurement, and the samples were outgassed at 120°C for 27 h under a reduced pressure of 6.58 × 10−5 Torr. X-ray photoelectron spectroscopy (XPS) spectra were obtained on a VG Thermo Escalab 250Xi spectrometer operated at 120 W. VSM analysis was carried out on the condition of hysteresis loop, which normal temperature and the maximum magnetic field applied were 2T. Temperature scanning rate was 1 K/min, vibration amplitude was 4 cm, and scanning twice each time. The uranyl ion concentration was determined by a trace uranium analyzer (WGJ-III, Institute of comparative geology of nuclear industry).

2.3. Synthesis of the Adsorption Materials

2.3.1. Synthesis of CuFe2O4 Nanoparticles. CuFe2O4 nanoparticles were prepared by the coprecipitation technique. 4.05 g FeCl3·6H2O FeCl3 and 2.93 g CuSO4·5H2O were dissolved in 200 mL 1 wt% PEG solution. Stir the solution continuously for about 60 minutes to ensure a balance between all
ingredients. The pH of the mixture was adjusted to 8 with adding 3 mol/L KOH solution. The mixture continued to be stirred magnetically for three hours and then aged overnight. The precipitate was filtered and washed repeatedly with distilled water until there were no SO$_4^{2-}$ and Cl$^{-}$ ions. Then, it was dried in a constant temperature drying oven at 50°C for 12 h and calcined at 600°C for 4 h, and finally, CuFe$_2$O$_4$ nanoparticles were obtained in the end.

2.3.2. Synthesis of CuFe$_2$O$_4$/PANI Nanocomposite. CuFe$_2$O$_4$/PANI nanocomposite was prepared by chemical in situ polymerization. 2 g CuFe$_2$O$_4$ nanoparticles were dispersed in 200 mL 2 mol/L hydrochloric acid solution and stirred vigorously for 10 minutes. A certain amount of 1.2 g distilled aniline monomers was added to the above mixture and then stirred continuously for 60 min. Then, 20 mL 20 mmol (NH$_4$)$_2$S$_2$O$_8$ solution was added drop by drop to the mixture as a polymerization initiator. An immediate change in the color of the molecules was observed. The suspension was stirred to complete the polymerization process in about 12 h.

Finally, the CuFe$_2$O$_4$/PANI nanocomposite was filtered, rinsed repeatedly with distilled water and anhydrous ethanol, and dried in a constant temperature drying oven at 70°C for 12 h.

2.4. Batch Adsorption Experiments. A series of uranium solutions with different concentrations (5 ~ 300 mg/L) were prepared. For adsorption studies, 5 mg CuFe$_2$O$_4$ nanoparticles and CuFe$_2$O$_4$/PANI nanocomposite was mixed, respectively, with 9 mL uranium solution with a concentration of 100 mg/L in a 10 mL centrifuge tube. The pH value of uranium solution was adjusted to 2 ~ 6 by with 1 mg/L hydrochloric acid and 0.1 mg/L sodium hydroxide, and the solution was adsorbed at 25°C, 35°C, and 45°C, respectively, for 5 min to 120 min. The mixture was centrifuged to take out the supernatant after absolutely adsorption, without inhaling the solid particles for intermittent adsorption experiment. Adsorption capacity ($Q_e$) is calculated according to Equation (1).
Figure 3: X-ray photoelectron spectroscopy of (a) CuFe$_2$O$_4$ nanoparticles and (b) CuFe$_2$O$_4$/PANI nanocomposite.
\[ Q_e = (C_0 - C_e) \times \frac{v}{m} \]  

(1)

In the above formula, \( C_0 \) and \( C_e \) were the initial concentration and equilibrium concentration of uranium solution (mg/L); \( v \) was the volume of uranium solution (L), \( m \) was the mass of adsorption material, and the unit was g.

2.5. Desorption Experiments. At the end of each adsorption experiment, the solution in the centrifuge tube was centrifuged to recover the adsorption material, which had adsorbed uranyl ions. The CuFe₂O₄/PANI nanocomposite was treated with ultrasound in 0.01 mol/L hydrochloric acid solution for 5 min to remove the uranyl ions adsorbed in the adsorption material. The uranyl ions in the solution were then separated from CuFe₂O₄/PANI nanocomposite by centrifugation. The recovered nanocomposite was washed thoroughly before the adsorption experiment. In the repeated adsorption experiment, the concentration of uranyl ions in the residual solution after adsorption was determined by the micro uranium analyzer. In order to further determine the recyclability of CuFe₂O₄/PANI nanocomposite, five adsorption/elution cycles of uranyl ions were repeated in the experiments [44].

2.6. Adsorption Kinetics. Adsorption kinetics was an important method to study the adsorption efficiency of uranium. In order to further study the adsorption kinetics and mechanism of CuFe₂O₄/PANI nanocomposite, the pseudosecond-order kinetic Equation (2) was used to fit the experimental data.

\[ \frac{t}{Q_t} = \frac{1}{K_2 \cdot Q_e^2} + \frac{t}{Q_e}. \]  

(2)

In the above formula, \( Q_e \) was the adsorption capacity of adsorption material for uranyl ions at equilibrium time (mg/g); \( Q_t \) was the adsorption capacity of the nanocomposite for uranyl ions reaction \( t \) time (mg/g); \( K_2 \) was the rate constant of quasi second order kinetics (g·mg⁻¹·min⁻¹).

2.7. Equilibrium Isotherm Modelling. In order to better study the adsorption mechanism of the adsorption materials for uranyl ions, Langmuir and Freundlich isothermal adsorption models were used to fit the experimental data for uranium adsorption. Langmuir and Freundlich isothermal adsorption equations were shown in Formulas (3) and (4), respectively.

\[ \frac{C_e}{q_e} = \frac{1}{q_{\text{max}}K_L} + \frac{C_e}{q_{\text{max}}}, \]  

(3)

\[ \ln Q_e = \ln K_f + \frac{n}{n} \ln C_e. \]  

(4)

In the above formula, \( C_e \) was the equilibrium concentration (mg/L); \( q_e \) was the adsorption capacity (mg/g) of
equilibrium time of uranyl ions; \( q_{\text{max}} \) was the maximum adsorption capacity (mg/g); \( K_L \) was the adsorption isotherm parameter of Langmuir; \( K_f \) represented adsorption strength (mg \( 1^{-n} \cdot g^{-1} \cdot L^n \)); \( N \) was the adsorption capacity.

3. Results and Discussion

3.1. Characterization. The TEM images of magnetic CuFe\(_2\)O\(_4\) nanoparticles and CuFe\(_2\)O\(_4\)/PANI nanocomposite are shown in Figure 1. As can be seen from the TEM images in Figure 1(a), CuFe\(_2\)O\(_4\) nanoparticles have irregular spherical shape with good dispersion and a particle size of about 100 nm. As shown in Figure 1(b), the generated polyaniline (PANI) coats with CuFe\(_2\)O\(_4\) magnetic nanoparticles and forms a core-shell structure. The boundary between CuFe\(_2\)O\(_4\)/PANI nanocomposite is clear and presents regular distribution.

It shows in Figure 2 that the XRD pattern of magnetic CuFe\(_2\)O\(_4\) nanoparticles and CuFe\(_2\)O\(_4\)/PANI nanocomposite are set from 10° to 90°. According to the XRD pattern, the diffraction peak is consistent with the amorphous phase of polyaniline, and the characteristic peak of PANI is displayed at 24.2°. It can be seen from the XRD pattern of CuFe\(_2\)O\(_4\) that the characteristic peak appears at 33.3°, and the diffraction peak is consistent with the square spinel of CuFe\(_2\)O\(_4\), with good crystallinity [45, 46]. When PANI is successfully combined with CuFe\(_2\)O\(_4\), it has no great influence on the crystal structure of CuFe\(_2\)O\(_4\) nanoparticles.

As shown in Figure 3(a), there are five peaks at 288.5, 399.1, 531.9, 710.7, and 950.2 eV, corresponding to C1s, N1s, O1s, Fe2p, and Cu2p, respectively. As shown in Figure 3(b), there are four peaks at 284.4, 532.1, 714.0, and 935.7 eV, corresponding to C1s, O1s, Fe2p, and Cu2p, respectively. It can be seen from the figure that a new peak of N1s with a binding energy of 400.7 eV appears, which proves the existence of nitrogen atoms. It indicates that polyaniline has been successfully compounded on the magnetic CuFe\(_2\)O\(_4\) nanoparticles.

Figure 4 shows the total pore volume of magnetic CuFe\(_2\)O\(_4\) nanoparticles and CuFe\(_2\)O\(_4\)/PANI nanocomposite of the N\(_2\) adsorption-desorption isotherm, which is shown in Table 1 for relevant parameters. Compared with CuFe\(_2\)O\(_4\)/PANI nanocomposites, the surface area of the material modified by polyaniline changes from 91.1 to 153 m\(^2\)/g, and the total pore volume and nitrogen adsorption capacity increase as the introduction of polyaniline. The porous core-shell structure formed by polyaniline coating on CuFe\(_2\)O\(_4\) increases the pore diameter of the nanocomposite.

It shows in Figure 5 the IR spectra of magnetic CuFe\(_2\)O\(_4\) nanoparticles and CuFe\(_2\)O\(_4\)/PANI nanocomposite. According to the IR of CuFe\(_2\)O\(_4\) nanoparticles, the peak at 433 cm\(^{-1}\) is the stretching vibration peak of Cu-O bond, and the peak at 597 cm\(^{-1}\) is the stretching vibration peak of Fe-O bond. This indicates that CuFe\(_2\)O\(_4\) nanoparticles have been successfully synthesized. According to the spectrum of CuFe\(_2\)O\(_4\)/PANI nanocomposite, the peak at 1301 cm\(^{-1}\) represents C-N vibration absorption peak, the peak at 1481 cm\(^{-1}\) is C-N vibration absorption peak, and the peak at 1567 cm\(^{-1}\) is N-H vibration absorption peak ([47–49]; Jawad, [50]). In the IR spectrum of CuFe\(_2\)O\(_4\)/PANI
Figure 6: VSM curves of CuFe$_2$O$_4$ nanoparticles, CuFe$_2$O$_4$/PANI nanocomposite, and adsorbed CuFe$_2$O$_4$/PANI nanocomposite.

Figure 7: Effect of initial pH value of solution on the adsorption of UO$_2$$^{2+}$ by CuFe$_2$O$_4$ nanoparticles and CuFe$_2$O$_4$/PANI nanocomposite.
Figure 8: Effect of initial concentration value of solution on the adsorption of $\text{UO}_2^{2+}$ by CuFe$_2$O$_4$ nanoparticles and CuFe$_2$O$_4$/PANI nanocomposite.

Figure 9: Effect of temperature and adsorption time value of solution on the adsorption of uranyl ions by CuFe$_2$O$_4$/PANI nanocomposite.
nanocomposite, the characteristic peak of PANI can be seen, indicating the successful preparation of CuFe\(_2\)O\(_4\)/PANI nanocomposite.

The magnetic properties of CuFe\(_2\)O\(_4\) nanoparticles, CuFe\(_2\)O\(_4\)/PANI nanocomposite, and adsorbed CuFe\(_2\)O\(_4\)/PANI nanocomposite are characterized by vibrating sample magnetometer (VSM). The VSM diagram of them is shown in Figure 6.

As can be seen from Figure 6, the synthesized CuFe\(_2\)O\(_4\) has magnetic properties (0.04 emu/g), but it is indeed very weak compared with other magnetic materials. It still appears "S" shaped hysteresis curve in CuFe\(_2\)O\(_4\), which could be simply separated from a solution by external magnetic fields. After coating with PANI, the saturation magnetization value seems increased to a certain extent, but in fact, the actual value is very low, only as 0.07 emu/g. The slight increase may be due to the residual free radicals during the polymerization process. After adsorption, the saturated magnetization intensity of CuFe\(_2\)O\(_4\)@PANI nanocomposite is reduced to 0.02 emu/g, which may be caused by the quenching of free radicals. It is close to its “intrinsic” value (0.025 emu/g, 0.04 emu/g \times 0.625, the ratio of CuFe\(_2\)O\(_4\) to PANI is 0.625).

3.2. Effect of pH on Adsorption of Uranyl Ions. Uranium has relatively active chemical properties. In alkaline media, uranium acid (UO\(_2^{2-}\)) or diuranic acid (U\(_2\)O\(_7^{2-}\)) anions are easily produced. Uranium grinding cations (UO\(_2^{2+}\)) tend to be produced in neutral and acidic media. The pH value of the solution not only affects the content of uranyl ions but also affects the adsorption performance of the adsorption material. Under the conditions of initial concentration of 100 mol/L, the adsorption time of 60 min and temperature of 25°C, CuFe\(_2\)O\(_4\)/PANI nanocomposite was used to study the effect of initial pH of the solution on the adsorption capacity of uranyl ions, shown as in Figure 7.

As can be seen from Figure 6, the synthesized CuFe\(_2\)O\(_4\) has magnetic properties (0.04 emu/g), but it is indeed very weak compared with other magnetic materials. It still appears "S" shaped hysteresis curve in CuFe\(_2\)O\(_4\), which could be simply separated from a solution by external magnetic fields. After coating with PANI, the saturation magnetization value seems increased to a certain extent, but in fact, the actual value is very low, only as 0.07 emu/g. The slight increase may be due to the residual free radicals during the polymerization process. After adsorption, the saturated magnetization intensity of CuFe\(_2\)O\(_4\)@PANI nanocomposite is reduced to 0.02 emu/g, which may be caused by the quenching of free radicals. It is close to its “intrinsic” value (0.025 emu/g, 0.04 emu/g \times 0.625, the ratio of CuFe\(_2\)O\(_4\) to PANI is 0.625).

In the initial solution, pH affects the adsorption experiment and the choice of pH range of 2~6 in the uranium adsorption experiments, and it can be seen that the adsorption quantity of magnetic CuFe\(_2\)O\(_4\) nanoparticles and CuFe\(_2\)O\(_4\)/PANI nanocomposite increases with the increase of pH value. At the condition of low pH, the adsorption capacity of CuFe\(_2\)O\(_4\)/PANI nanocomposite as well as magnetic CuFe\(_2\)O\(_4\) nanoparticles is low, and it is because there is a lot of H\(^+\) that compete with uranyl ion adsorption. With the increase of pH, the adsorption rate increases. If the pH value continues to increase, at a higher pH value, UO\(_2^{2+}\) will hydrolyze to produce UO\(_2\)(OH)\(_2\) polynuclear hydroxide precipitation. So the pH value of 4 is selected for the following adsorption experiment.

As shown in Figure 7, the adsorption performance of CuFe\(_2\)O\(_4\)/PANI nanocomposite is higher than that of CuFe\(_2\)O\(_4\) when the pH value is very low. When the pH value is in the range of 4 to 6, the adsorption performance gradually decreases.
3.3. Effect of Initial Concentration on Adsorption of Uranyl Ions. In order to evaluate the adsorbability of the materials and further to prove whether the introduction of amino groups could improve the adsorption properties. As can be seen from Figure 8, the adsorption capacity of CuFe₂O₄/PANI nanocomposite and CuFe₂O₄ nanoparticles is relatively small when the initial concentration of uranyl ions is low, and the adsorption capacity goes up with the increase of the initial concentration. When the initial concentration is low, more amino groups would be coordinated with uranyl ions, and up to saturation point as the initial concentration increases gradually. So, the adsorption rate of uranyl ions in CuFe₂O₄/PANI nanocomposite is relatively high at low concentration. As a large number of nitrogen exist in CuFe₂O₄/PANI nanocomposite, the adsorption amount is as high as almost 320 mg/g under the condition of Ce < 100 mg/L in its quick equilibrium state which indicates the excellent adsorption performance for uranyl ions adsorption.

3.4. Effect of Temperature and Adsorption Time on Adsorption of Uranyl Ions. It shows the adsorption capacity curve of the adsorption material at temperatures of 25°C, 35°C, and 45°C with the adsorption time of 5 ~ 120 min in Figure 9. It can be observed that the adsorption capacity of the adsorbent is basically balanced at about 20 min, and the adsorption equilibrium time is short. It can be seen from the results that the adsorption capacity of the adsorption materials is affected by the temperature. CuFe₂O₄/PANI nanocomposite adsorbing uranyl ions is the heat absorption process, the adsorption capacity is significantly lower when the temperature is low, and the adsorption capacity is greater when the temperature is high. This is mainly because with the

![Figure 11: Pseudosecond-order kinetic curves of uranium adsorption on CuFe₂O₄/PANI nanocomposite at different temperatures: (A) 25°C, (B) 35°C, and (C) 45°C, respectively.](image)

| T (°C) | qₑ (mg/g) | K₂ (g·mg⁻¹·min⁻¹) | R²   |
|-------|----------|------------------|------|
| 25    | 131.6    | 0.0064           | 0.9975|
| 35    | 163.9    | 0.0081           | 0.9994|
| 45    | 188.7    | 0.0055           | 0.9993|
increase of adsorption temperature, the reaction activity of amino groups and uranyl ions of CuFe$_2$O$_4$/PANI nanocomposite is enhanced, which is conducive to display its adsorption performance.

3.5. Regeneration. Recycling is an important index to evaluate the adsorption and regeneration performance of the adsorbents. It is of great significance to evaluate recycling in the practical application of adsorbents, which can not only save time but also reduce production costs. In order to evaluate the important index of recycling regeneration of CuFe$_2$O$_4$/PANI nanocomposite, 0.01 mol/L HCl solution was used as the elution solution of regeneration.

More significantly, as reflected in Figure 10, the magnetic adsorption material is used to remove uranyl ions after regeneration. It shows the adsorption rate of CuFe$_2$O$_4$/PANI nanocomposite recovered in each adsorption/desorption cycle. The adsorption capacity of uranyl ions displays more than 170 mg/g during five consecutive cycles. The slight influence on the adsorption capacity proves that the CuFe$_2$O$_4$/PANI nanocomposite has a high reuse rate and displays its practical application prospect.

3.6. Adsorption Kinetics. Adsorption kinetics is an important method to study the adsorption efficiency of uranium. It can be seen from Figure 11 that CuFe$_2$O$_4$/PANI nanocomposite exhibits fast adsorption kinetics for uranyl ions.

It can be seen from Figure 11 and Table 2 that both adsorption capacity and reaction rate are affected by temperature. The quasisecond-order kinetic model to calculate the actual value theory adsorption capacity is closed to the experiment, and it shows that quasisecond-order kinetic model fits the linear correlation coefficient. The correlation coefficient $R^2$ is more than 0.99; therefore, the quasisecond-order kinetic model can describe the adsorption behavior of CuFe$_2$O$_4$/PANI nanocomposite for uranyl ions.

3.7. Equilibrium Isotherm Modelling. In order to better study the adsorption mechanism of CuFe$_2$O$_4$ and CuFe$_2$O$_4$/PANI nanocomposite for uranyl ions, Langmuir and Freundlich isothermal adsorption models are used to fit the experimental data for uranyl ions adsorption of CuFe$_2$O$_4$ and CuFe$_2$O$_4$/PANI nanocomposite, respectively.

As shown in Figures 12 and 13 and Table 3, the correlation coefficient $R^2$ of the Langmuir model is greater than 0.999. The experimental data are closed to the theoretical model, indicating that it can well describe the adsorption process of uranyl ions by CuFe$_2$O$_4$/PANI nanocomposite. The isotherm is best described by Langmuir adsorption model in the process of adsorption kinetics.

The maximum adsorption capacity for uranyl ions is about 322.6 mg/g, which is higher than many other types of uranyl ion adsorbents and has good adsorption performance comparing with other similar work [13].

It lists the comparison of maximum adsorption capacity ($q_m$) of U(VI) using various adsorbents in Table 4 [49, 51–...
The maximum adsorption capacity of CuFe$_2$O$_4$/PANI on U(VI) was higher than that of other adsorbents as shown in Table 4.

#### Table 3: Langmuir and Freundlich isotherm adsorption constants of CuFe$_2$O$_4$ nanoparticles and CuFe$_2$O$_4$/PANI nanocomposite for uranium adsorption.

| Adsorbent type | $K_L$ (mg/L) | $q_m$ (mg/g) | $R^2$ | $K_f$ (mg$^{1-n}$·g$^{-1}$·L$^n$) | $n$ | $R^2$ |
|----------------|---------------|---------------|-------|---------------------------------|-----|-------|
| CuFe$_2$O$_4$  | 0.0709        | 49.50         | 0.9996| 5.57                            | 2.38 | 0.8992|
| CuFe$_2$O$_4$/PANI | 0.310         | 322.6         | 0.9998| 102                             | 3.74 | 0.8545|

#### Table 4: Comparison of adsorption capacity of CuFe$_2$O$_4$/PANI nanocomposite with different adsorbents.

| Adsorbents | Maximum adsorption capacity | Ref |
|------------|----------------------------|-----|
| Polypropylene (PP) nanofibers | 83.24 mg/g | [55]|
| Ap-ZnO/PSBN | 139.8 mg/g | [51]|
| AM/PDA | 250.7 mg/g | [54]|
| MoS$_2$-IP6 NRA/CC | 183.3 mg/g | [53]|
| SBA-15/NH$_2$SO$_4$H | 140.5 mg/g | [52]|
| CuFe$_2$O$_4$/PANI | 322.6 mg/g | This work|

### 3.8. Magnetic Validation Study.

As shown in Figure 13, the magnetic verification experiment results show that the synthesized CuFe$_2$O$_4$/PANI nanocomposite has good magnetic properties, which ensures that the adsorption material can be easily separated from the cleaning medium. It shows in the magnetic verification experiment the ideal magnetism under the action of magnet in Figure 14(a). On this basis, its dispersion is further verified by ultrasound. Results as shown in Figure 14(b), CuFe$_2$O$_4$/PANI magnetic nanocomposite has good dispersion after ultrasound, indicating that the particles are uniform and of good particle size. In the magnetic separation verification experiment of SiO$_2$ turbidity liquid, it has sufficient magnetic response to make it separated from the mixture immediately under the action of the external magnetic field. Therefore, CuFe$_2$O$_4$/PANI nanocomposite has been easily to be enriched under the action of magnetic field and can be used for magnetic separation, recovery, and reuse for its practical application prospect.
4. Conclusion

In this study, a novel magnetic adsorption material CuFe$_2$O$_4$/PANI nanocomposite has been successfully prepared by polyaniline (PANI) combined with CuFe$_2$O$_4$ nanoparticles in situ polymerization and is used for the adsorption of radioactive uranium in an aqueous solution. When pH = 4.0, magnetic CuFe$_2$O$_4$/PANI nanocomposite has the best adsorption effect, and the maximum adsorption capacity for uranyl ions is 322.6 mg/g. The introduction of amino functional groups makes the adsorption performance of CuFe$_2$O$_4$/PANI nanoparticles much higher than that of CuFe$_2$O$_4$ nanoparticles. The adsorption process is consistent with the quasisecond-order kinetic equation and Langmuir
isothermal adsorption model, which indicates that the adsorption process is mainly chemisorption and monolayer adsorption. The recovery experiments show that magnetic CuFe$_2$O$_4$/PANI nanocomposite has good repeatability and stability due to its magnetic properties. After five cycles of recovery, the adsorption capacity of magnetic CuFe$_2$O$_4$/PANI nanocomposite would be maintained at more than 170 mg/g, which is expected to be a promising renewable adsorption material.

**Data Availability**

The datasets generated during and/or analysed during the current study are available from the corresponding author on reasonable request.

**Additional Points**

**Highlights.** A novel magnetic adsorption material CuFe$_2$O$_4$/PANI nanocomposite has been synthesized by polyaniline (PANI) combined with CuFe$_2$O$_4$ in situ polymerization. The prepared CuFe$_2$O$_4$/PANI nanocomposite exhibits extremely high maximum adsorption capacity (322.6 mg/g) for uranyl ion in wastewater at a pH of 4 and can quickly reach the equilibrium state. The magnetic CuFe$_2$O$_4$/PANI displays stable adsorption performance for uranyl ions after five cycles of recovery, which indicates it possesses good recovery due to its magnetism and excellent regeneration ability for reusability.

**Conflicts of Interest**

The authors declare that they have no conflicts of interest.

**References**

[1] M. J. Manos and M. G. Kanatzidis, “Layered metal sulfides capture uranium from seawater,” *Journal of the American Chemical Society*, vol. 134, no. 39, pp. 16441–16446, 2012.

[2] S. Chu and A. Majumdar, “Opportunities and challenges for a sustainable energy future,” *Nature*, vol. 488, no. 7411, pp. 294–303, 2012.

[3] R. L. Kathren and S. Y. Tolmachev, “The US Transuranium and Uranium Registries (USTUR): a five-decade follow-up of plutonium and uranium workers,” *Health Physics*, vol. 117, no. 2, pp. 118–132, 2019.

[4] D. Ma, J. Wei, Y. Zhao, Y. Chen, and S. Tang, “The removal of uranium using novel temperature sensitive urea-formaldehyde resin: adsorption and fast regeneration,” *Science of the Total Environment*, vol. 735, p. 139399, 2020.

[5] T. Özdemir and A. Usanmaz, “Use of poly(methyl methacrylate) in radioactive waste management: II. Monte Carlo simulations,” *Progress in Nuclear Energy*, vol. 51, no. 8, pp. 845–848, 2009.

[6] A. Mellah, S. Chegrouche, and M. Barkat, “The precipitation of ammonium uranyl carbonate (AUC): thermodynamic and kinetic investigations,” *Hydrometallurgy*, vol. 85, no. 2-4, pp. 163–171, 2007.

[7] M. M. Chen, Z. Li, J. H. Li, J. Li, Q. Li, and L. Zhang, “The extraction of uranium using graphene aerogel loading organic solution,” *Talanta*, vol. 166, pp. 284–291, 2017.

[8] X. Y. Zheng, X. Y. Wang, Y. H. Shen, X. Lu, and T. S. Wang, “Biosorption and biomineralization of uranium(VI) by Sarccharomyces cerevisiae – crystal formation of chernikovite,” *Chemosphere*, vol. 175, pp. 161–169, 2017.

[9] M. G. Torkabadi, A. R. Keshkar, and S. J. Salfdari, “Comparison of polylethylene sulfone and polyaamide nanofiltration membranes for uranium removal from aqueous solution,” *Progress in Nuclear Energy*, vol. 94, pp. 93–100, 2017.

[10] J. De Decker, K. Folens, J. De Clercq et al., “Ship-in-a-bottle CMPO in MIL-101(Cr) for selective uranium recovery from aqueous streams through adsorption,” *Journal of Hazardous Materials*, vol. 335, pp. 1–9, 2017.

[11] S. A. Kirillov, T. V. Lisnycha, and O. I. Pendelyuk, “Appraisal of mixed amorphous manganese oxide/titanium oxide sorbents for the removal of strontium-90 from solutions, with special reference to Savannah river site and Chernobyl radioactive waste simulant,” *Adsorption Science & Technology*, vol. 24, no. 10, pp. 895–906, 2006.

[12] F. Liu, S. Song, G. Cheng, W. Xiong, L. Shi, and Y. Zhang, “MIL-101(Cr) metal-organic framework functionalized with tetraethylenglenepentamine for potential removal of uranium (VI) from wastewater,” *Adsorption Science & Technology*, vol. 36, no. 7-8, pp. 1550–1567, 2018.

[13] A. A. Alqadami, M. Naushad, Z. A. Alothaman, and A. A. Ghfar, “Novel metal-organic framework (MOF) based composite material for the sequestration of U(VI) and Th(IV) metal ions from aqueous environment,” *ACS Applied Materials & Interfaces*, vol. 9, no. 41, pp. 36026–36037, 2017.

[14] M. d’Halluin, J. Rull-Barrull, G. Bretel, C. Labruguère, E. Le Grognec, and F.-X. Felpin, “Chemically modified cellulose filter paper for heavy metal remediation in water,” *ACS Sustainable Chemistry & Engineering*, vol. 5, no. 2, pp. 1965–1973, 2017.

[15] A. Z. Ekoa Bessa, G. Ngueuchoua, A. Kwewou Janpow et al., “Heavy metal contamination and its ecological risks in the beach sediments along the Atlantic Ocean (Limb coastal fringes, Cameroon),” *Earth Systems and Environment*, pp. 1–12, 2020.

[16] S. Rangabhashiyam and P. Balasubramanian, “Characteristics, performances, equilibrium and kinetic modeling aspects of heavy metal removal using algae,” *Bioresource Technology Reports*, vol. 5, pp. 261–279, 2019.

[17] G. Sharma and M. Naushad, “Adsorptive removal of noxious cadmium ions from aqueous medium using activated carbon/zirconium oxide composite: isotherm and kinetic modelling,” *Journal of Molecular Liquids*, vol. 310, pp. 318–327, 2020.

[18] X. L. Liu, H. W. Pang, X. W. Liu et al., “Orderly porous covalent organic frameworks-based materials: superior adsorbents for pollutants removal from aqueous solutions,” *The Innovation*, vol. 2, no. 1, p. 100076, 2021.

[19] A. Zhang, G. Uchiyama, and T. Asakura, “Dynamic-state adsorption and elution behaviour of uranium (VI) ions from seawater by a fibrous and porous adsorbent containing amidoxime chelating functional groups,” *Adsorption Science & Technology*, vol. 21, no. 8, pp. 761–773, 2003.

[20] W. H. Zou, J. Y. Song, K. Li, and R. Han, “Use of manganese oxide-coated sand for the adsorption of uranium(VI) ions from aqueous solution using a column mode,” *Adsorption Science & Technology*, vol. 28, no. 4, pp. 313–325, 2010.

[21] W. Wongjaiakh, D. Wongsawang, P. Hosemann, C. Kanokworakan, and V. Ratnitsai, “Enhancement of uranium recovery from seawater using amidoximated polymer
gel synthesized from radiation-polymerization and crosslinking of acrylonitrile and methacrylic acid monomers,” *Journal of Environmental Chemical Engineering*, vol. 6, no. 2, pp. 2768–2777, 2018.

[22] J. L. Wu, K. Tian, and J. L. Wang, “Adsorption of uranium (VI) by amidoamine modified multilayered carbon nanotubes,” *Progress in Nuclear Energy*, vol. 106, pp. 79–86, 2018.

[23] P. Zong, X. Wu, J. Gou, X. Lei, and H. Deng, “Immobilization and recovery of uranium(VI) using Na-Bentonite from aqueous medium: equilibrium, kinetics and thermodynamics studies,” *Journal of Molecular Liquids*, vol. 209, pp. 358–366, 2015.

[24] T. S. Anirudhan, G. S. Lekshmi, and F. Shainy, “Characteristics of equilibrium, kinetics studies for adsorption of Hg(II) and Cr(VI) by polyaniline/humic acid composite,” *Desalination*, vol. 266, no. 1-3, pp. 188–194, 2011.

[25] M. Bhauamik, A. Maity, V. V. Srinivasu, and M. S. Onyango, “Removal of hexavalent chromium from aqueous solution using polypropylene-polyaniline nanofibers,” *Chemical Engineering Journal*, vol. 181, pp. 323–333, 2012.

[26] D. D. Shao, C. L. Chen, and X. K. Wang, “Application of polyaniline and multiwalled carbon nanotube magnetic composites for removal of Pb (II),” *Chemical Engineering Journal*, vol. 185, pp. 144–150, 2012.

[27] M. Ben Ali, F. Y. Wang, B. Bokherroub, W. Lei, and M. Xia, “Phytic acid-doped polyaniline nanofibres-clay mineral for efficient adsorption of copper (II) ions,” *Journal of Colloid and Interface Science*, vol. 553, pp. 688–698, 2019.

[28] R. B. Gapusan and M. D. L. Bulela, “Adsorption of anionic methyl orange dye and lead(II) heavy metal ion by polyaniline-kapok fiber nanocomposite,” *Materials Chemistry and Physics*, vol. 243, p. 122682, 2020.

[29] S. A. Hashemi, S. M. Mousavi, and S. Ramakrishna, “Effective removal of mercury, arsenic and lead from aqueous media using polyaniline-Fe3O4–silver diethyldithiocarbamate nanostructures,” *Journal of Cleaner Production*, vol. 239, p. 118023, 2019.

[30] H. Saghatchi and R. Ansari, “Application of magnetic polyaniline nanocomposite for separation of uranyl ions from aqueous solutions,” *Separation Science and Technology*, vol. 53, no. 15, pp. 2486–2499, 2018.

[31] D. D. Shao, G. S. Hou, J. X. Li, T. Wen, X. Ren, and X. Wang, “PANI/GO as a super adsorbent for the selective adsorption of uranium(VI),” *Chemical Engineering Journal*, vol. 255, pp. 604–612, 2014.

[32] B. Li, L. J. Ma, Y. Tian et al., “A catechol-like phenolic ligand-functionalized hydrothermal carbon: one-pot synthesis, characterization and sorption behavior toward uranium,” *Journal of Hazardous Materials*, vol. 271, pp. 41–49, 2014.

[33] M. Gholinejad, J. Ahmadi, and C. Najera, “Silica microparticles supported gold and copper ferrite nanoparticles: a magnetically recyclable bimetallic catalyst for Sonogashira reaction,” *ChemistrySelect*, vol. 1, no. 3, pp. 384–390, 2016.

[34] J. Kuriand M. J. Mathew, “Structural, optical and magnetic studies of CuFe2O4, MgFe2O4 and ZnFe2O4 nanoparticles prepared by hydrothermal/ solvothermal method,” *Journal of Magnetism and Magnetic Materials*, vol. 451, pp. 121–130, 2018.

[35] A. H. Jawad and A. S. Abdulhameed, “Facile synthesis of cross-linked chitosan-tripolyphosphate/kaolin clay composite for decolourisation and COD reduction of remazol brilliant blue R dye: optimization by using response surface methodology,” *Colloids and Surfaces A: Physicochemical and Engineering Aspects*, vol. 605, article 125329, 2020.

[36] Z. D. Zuzovic, Y. Wang, G. A. Bowmaker, and R. B. Kaner, “Structure of ultralong polyaniline nanofibers using initiators,” *Macromolecules*, vol. 44, no. 8, pp. 2735–2742, 2011.

[37] N. N. A. Malek, A. H. Jawad, A. S. Abdulhameed, K. Ismail, and B. H. Hameed, “New magnetic Schiffl’s base-chitosan-
glyoxal/fly ash/Fe₃O₄ biocomposite for the removal of anionic azo dye: an optimized process,” *International Journal of Biological Macromolecules*, vol. 146, pp. 530–539, 2020.

[50] A. H. Jawad, N. S. A. Mubarak, and A. S. Abdulhameed, “Tunable Schiff’s base-cross-linked chitosan composite for the removal of reactive red 120 dye: adsorption and mechanism study,” *International journal of biological macromolecules*, vol. 142, pp. 732–741, 2020.

[51] M. F. Cheira, “Synthesis of aminophosphonate-functionalised ZnO/polystyrene-butadiene nanocomposite and its characteristics for uranium adsorption from phosphoric acid,” *International Journal of Environmental Analytical Chemistry*, pp. 1–25, 2019.

[52] Z. Dousti, L. Dolatyari, M. R. Yaftian, and S. Rostamnia, "Adsorption of Eu(III), Th(IV), and U(VI) by mesoporous solid materials bearing sulfonic acid and sulfamic acid functionalities," *Separation Science and Technology*, vol. 54, no. 16, pp. 2609–2624, 2019.

[53] Y. Kou, L. Zhang, B. Liu, L. Zhu, and T. Duan, "Phosphonate modified MoS₂ composite material for effective adsorption of uranium(VI) in aqueous solution," *Journal of Radioanalytical and Nuclear Chemistry*, vol. 323, no. 1, pp. 641–649, 2020.

[54] H. Yang, H. Ding, X. Zhang, X. Luo, and Y. Zhang, "Immobilization of dopamine on Aspergillus niger microspheres (AM/PDA) and its effect on the U(VI) adsorption capacity in aqueous solutions," *Colloids and Surfaces a-Physicochemical and Engineering Aspects*, vol. 583, article 123914, 2019.

[55] F. Ashrafi, M. Firouzzare, S. J. Ahmadi, M. Sohrabi, and M. Khosravi, "Preparation and modification of forcespun polypropylene nanofibers for adsorption of uranium (VI) from simulated seawater," *Ecotoxicology and environmental safety*, vol. 186, p. 109746, 2019.