Abstract: Urban natural surfaces and non-surface human activities are key factors determining the urban heat island (UHI), but their relative importance remains highly controversial and may vary at different spatial scales and focal urban systems. However, systematic studies on the scale-dependency system-specificity remain largely lacking. Here, we selected 32 major Chinese cities as cases and used Landsat 8 images to retrieve land surface temperature (LST) and quantify natural surface variables using point of interest (POI) data as a measure of the human activity variable and using multiple regression and relative weight analysis to study the contribution and relative importance of these factors to LST at a range of grain sizes (0.25–5 km) and spatial extents (20–60 km). We revealed that the contributions and relative importance of natural surfaces and human activities are largely scale-dependent and system-specific. Natural surfaces, especially vegetation cover, are often the most important UHI determinants for a majority of scales, but the importance of non-surface human activities is increasingly pronounced at a coarser spatial scale with respect to both grain and spatial extent. The scaling relations of the UHI determinants and their relative importance were mostly linear-like at the city-collective level, but highly diverse across individual cities, so reducing non-surface heat emissions could be the most effective measure in particular cases, especially at relatively large spatial scales. This study advances the understanding of UHI formation mechanisms and highlights the complexity of the scale issue underpinning the UHI effect.

Keywords: land surface temperature; urban heat island; natural surface; human activity; point of interest; multi-scale analysis; scaling

1. Introduction

According to the projection of the United Nations, urban residents will account for 70% of the world’s total population by the middle of this century [1]. Urbanization has profoundly transformed the Earth’s land surfaces and modified the composition, structure, and energy flow of the atmospheric system in the near-surface layer [2,3]. Among these remarkable urbanization-induced modifications is the urban heat island (UHI) effect, referred to as the phenomenon of urban temperatures being higher...
than those in rural surroundings [4–8]. Studies on the UHI effect are gaining momentum across a wide range of research fields, with particular concerns given to its negative environmental and social impacts such as deteriorated air quality [9], excessive consumption of water and power [2], and impaired human health [10]. Therefore, there is an urgent need for us to take measures to help mitigate the UHI effect. Furthermore, a comprehensive, quantitative understanding of the determinants of UHIs would help to choose the most efficient mitigating measures.

Typically, the UHI effect can be quantified based on the measurement of near-surface air temperature (referred to as atmospheric UHIs) or land surface temperature (referred to as surface UHIs) [11,12]. Facilitated by readily available remotely sensed data, recent studies have increasingly focused on land surface temperature (LST) as a direct indicator of the surface UHI effect to study how UHIs are shaped by an array of environmental and anthropogenic determinants [13–16]. It has been well documented that remotely sensed LST is strongly correlated with natural-surface biophysical variables that characterize vegetation and water cover, suggesting that urban vegetation and water bodies have a significant cooling effect [4,14,17,18]. A set of remotely sensed indicators of natural-surface properties such as the normalized difference vegetation index (NDVI), the leaf area index (LAI), and the normalized difference water index (NDWI) has therefore been adopted in studies aimed at unraveling the driving factors of UHIs [19,20].

However, these natural-surface factors are not the only determinants of UHIs [6] because some human activities that do not modify the physical properties of land surfaces can also contribute to UHI formation. For example, two separate land units that accommodate substantially different resident densities, with all else being equal, would emit different amounts of heat (for the purpose of e.g., stabilizing indoor environments), thus resulting in differential local thermal environments. Indeed, a handful of empirical studies has suggested that population density and societal status (in terms of income and education level) have significant effects driving the formation of UHIs [7,10,21].

One barrier for us in the study of the human influence of LST is that the quick development of remote sensing technology allows us to extract high-resolution LST and natural-surface data, while the social sensing data are usually restricted by the administrative census. Fortunately, the rapid development of Internet technology and the increasing availability of open big data are substantially alleviating this difficulty. In recent years, the development of social sensing technology has produced a rich set of high-resolution data, allowing for the quantitative characterization of human activities in detail [22,23]. Many social sensing data (such as social media data, public transport data, and location-based service data) have been mined, giving rise to a range of interesting socioeconomic structures and dynamics that have been previously impossible to reveal [22,24]. One particular useful source is the Point of Interest (POI) data, which provides location-based geographic information and other detailed information for each geographic element on the map [25]. POI contains almost all types of human activities in the city, and are suggested to be able to well reflect the intensity of human activities and anthropogenic heat emissions [22,25,26]. While increasing attention is being paid to these non-surface factors, their strengths relative to the land-surface factors remain unclear.

Assessing the relative importance of the surface vs. non-surface factors in shaping the UHI effect (hereafter referred to as the relative importance problem, RIP) is highly relevant to the mechanistic understanding of the UHI effect. It also has important implications for the decision making of urban climate mitigation [27,28]. For example, at specific conditions, if a UHI pattern is predominantly determined by the natural surface properties, the effectiveness of measures for cooling down the urban environment would largely rely on the design of the urban green system. However, the answer to the RIP remains elusive, as contrasting conclusions have been reached in previous studies. On one hand, it has been suggested that compared with human activities, the biophysical properties of land (especially natural) surfaces can largely determine the latent and sensible heat flux and generally contribute the most to LST pattern and intensity [2,3,29]. On the other hand, it has been stressed that non-surface human activities also play a key role in shaping UHIs through contributing an important, direct heat source, particularly in high-density neighborhoods at local spatial scales [5,6,8,30].
This active debate on the RIP of UHI determinants may be undermined by some important confounding factors that have not been explicitly addressed. First, it remains poorly understood if and to what extent the RIP answer is specific or universal to different urban systems. Regarding previous studies that have largely been based on single urban systems, a more comprehensive understanding would necessarily invoke systematic investigations that take into account multiple urban systems across a wide geographic range. Second, it remains unclear if and how the RIP answer is dependent on the observational scale. Previous work has shown that the spatial pattern of LST, as well as its driving factors, are sensitive to the spatial resolution of the remotely sensed data used [15,21,27,28,31], implying that the RIP solution is plausibly dependent on grain size (as a fundamental dimension of spatial scale). A complete understanding of the role of scale requires systematic scaling analyses that consider both the spatial extent and grain size.

Therefore, in this study, to clearly answer these insufficiencies, we assessed the relative strengths of natural-surface vs. non-surface factors in shaping the LST patterns across 32 major Chinese cities. We focused on the natural surfaces (including vegetation and water) because they have been recognized as essential "ecological infrastructure" for mitigating the UHI effect. We used the Landsat-8 OLI remote sensing data and the POI social sensing data to characterize natural-surface properties and non-surface human activities, respectively. We then assessed their relative importance in explaining the observed LST patterns across a range of spatial scales in terms of grain size and spatial extent. We aimed to answer two specific questions: (1) Is the relative importance between natural-surface vs. non-surface factors dependent on observational scales? and (2) Is the relative importance specific to different urban systems? Finally, we hope that this study will provide implications for understanding the mechanism of UHI formation and UHI mitigation practices.

2. Materials and Data Processing

2.1. Study Areas and Data Pre-Processing

For our study, we selected major Chinese cities with built-up areas >100 km$^2$ in 2015. For this selection, as a first step, we extracted all built-up areas within China using the global land cover product at a spatial resolution of 300 m from the Climate Change Initiative of the European Space Agency (ESACCI-LC-L4-LCCS-Map-300m-P1Y-2015-v2.0.7) [32]. By filtering out those <100 km$^2$, we yielded 70 candidate cities. The social sensing data were real-time data and we collected it in October 2017, so we checked all available cloud-free Landsat-8 OLI images acquired during 2015–2017 to ensure the time-efficiency of the data. Eventually, 32 cities that had available cloud-free images during this time period were defined as our study areas (Figure 1, see Table A1, Appendix A for detailed information). All Landsat-8 images used in this study were geometrically registered. For each image, we conducted radiometric calibration and then atmospheric correction using the FLAASH tool with the ENVI 5.2 software. The resulting reflectance images were used for subsequent analyses.
2.2. Land Surface Temperature Retrieval

We used the radiative transfer equation (RTE) algorithm [33] for the Landsat-8 thermal infrared sensor (TIRS) to quantitatively retrieve the LST in all studied cities, considering that the accuracy of this algorithm has been repeatedly validated [8,17]. A brief description of the RTE algorithm is given below:

\[ L_{\lambda} = [\varepsilon B(T_s) + (1-\varepsilon) L_d] \tau + L_u, \]  
\[ (1) \]

where \( L_{\lambda} \) is the apparent radiance received by the sensor; \( \varepsilon \) is the surface emissivity; \( T_s \) is the land surface temperature; \( B(T_s) \) is blackbody spectral radiance; \( \tau \) is atmospheric transmittance; \( L_u \) is the upwelling path radiance; and \( L_d \) is the downwelling path radiance. According to Plank’s Law, \( B(T_s) \) can be calculated as:

\[ B(T_s) = \frac{[L_{\lambda} - L_u - \tau(1-\varepsilon)L_d]}{\tau \varepsilon}, \]  
\[ (2) \]

Then, the LST \( T_s \) can be calculated as:

\[ T_s = K_2 / \ln(K_1 / B(T_s) + 1), \]  
\[ (3) \]
The atmospheric parameters $\tau$, $L_u$, and $L_d$ can be calculated with the Atmospheric Correction Parameter Calculator [34]. For Landsat-8 TIRS band 10, $K_1 = 774.89 \text{ mW m}^{-2} \text{s}^{-1} \text{sr}^{-1} \mu\text{m}^{-1}$; $K_2 = 1321.08 \text{ K}$.

The retrieved LST maps can be found in Appendix A (Figure A2).

2.3. Natural-Surface and Non-Surface Variables

Prior to analyses, the areas were divided into grid cells at a particular size (see Section 3.2 below). Within each grid cell, we calculated the natural-surface and non-surface variables. Our natural-surface variables included vegetation cover and water cover (seen in Figure A3, Appendix A) as they are the most commonly used variables in the studies of the surface UHI effect [4,14,19,35]. Within each grid cell, we calculated the mean of the normalized difference vegetation index (NDVI) for all vegetated pixels as an indicator of vegetation cover. For water cover, we calculated the percentage of water bodies within each grid cell.

We used POI as an indicator of the intensity of human activities. POI is a point type geo-located data representing all urban facilities such as companies, shopping malls, schools, and parks. Unlike traditional remote sensing data where one building is one point, POI data can give more detailed information within the building such as the shops, restaurants, beauty shops, etc. close by. More POI density means more human activity intensity. POI data are real-time data and we collected the data in October 2017. The POI data were obtained from the AMap [36], which is one of the most widely used web mapping and navigation services in China. We derived the spatial data of POIs for each studied city using the application programming interface (API) provided by AMap (Figure A4, Appendix A). We then calculated the density of POIs within each grid cell.

3. Methodology

3.1. Statistical Analyses

To assess the explanatory power of the natural-surface vs. non-surface variables, we constructed regression models with LST as the response variable and NDVI, water cover, and POI density as the explanatory variables (one grid cell represents one data point in the models). NDVI, water cover, and POI density were log-transformed to approach normal distributions before model fitting.

As a first step, we constructed the most commonly used univariate ordinary least squares (OLS) models to assess the explanatory power (measured by $R^2$) of each individual explanatory variable in explaining the observed LST patterns [28,37]. Then, we combined all three explanatory variables to fit multivariate OLS models to assess their collective explanatory power (measured by adjusted $R^2$). We used the relative importance analysis (RWA) to partition the relative importance (measured by the proportional contribution of the variables decomposed from the total explanatory power) of the three explanatory variables [38]. In addition, to account for the potential spatial autocorrelation (SAC) problem in the OLS models, we used simultaneous autoregressive (SAR) models to test for the robustness of our results. We used the spatial error model (SAR$\text{err}$) recommended by Kissling and Carl [39], considering the uncertainty of the autocorrelation sources [40,41]. The explanatory power of the SAR models was decomposed into spatial signal versus non-spatial trends. We subsequently calculated the pseudo-$R^2$ of the non-spatial component, and by checking the result of the SAR$\text{err}$ model for explaining the LST pattern (Figure A19, Appendix B), we found that our OLS scaling relations were qualitatively robust to spatial autocorrelation. Considering that a comparison with other studies where the OLS model was mostly used as well as the method for the application of the RWA to SAR models are not available yet, we only used the results of the OLS model to answer our questions in this article. The statistical analyses were performed with R 3.3.4. The RWA was performed by using the RWA-Web software [42].
3.2. Scaling Analysis

We conducted a straightforward scaling analysis by systematically varying both the spatial extent and grain size to quantify the individual explanatory power of NDVI, water cover, and POI density as well as their collective explanatory power at different spatial scales. In doing so, for each city, we restricted the study area within a squared region according to the urban areas, and then divided this squared region into grid cells; we changed the size of the squared region (i.e., 20 × 20 km², 30 × 30 km², 40 × 40 km², 50 × 50 km², and 60 × 60 km²) for the varying spatial extent, and changed the size of the within grid cells (i.e., 0.25 × 0.25 km², 0.5 × 0.5 km², 1 × 1 km², 2 × 2 km², and 5 × 5 km²) for varying grain size (see Figure A1, Appendix A for the schematic diagram of the scaling analysis). For the selection of the spatial grain size, 500 m and 1000 m are the most commonly used grain sizes for the study of the UHI effect and so we made the scaling ladder approximate a double geometric progression that could be divisible by the spatial extents. At each combination of spatial extent and grain size, we conducted all the statistical analyses (in Section 3.1) and the scaling relations were assessed by looking at how the explanatory power changed across the scale ranges.

The materials used, data processing, and the procedures carried out are shown in the flowchart below (Figure 2).

![Flowchart showing the procedures for data acquisition and methodology.](image)

4. Results

4.1. Scaling Relations at the City-Collective Level

The natural-surface (NDVI and water cover) and non-surface (POI density) variables jointly explained a considerable proportion (ca. 50–90%) of LST variance for all studied cities (Figure 3),
suggesting that they are indeed good predictors of the UHI patterns. Our scaling analysis clearly demonstrated that for the selected variables, both separately and collectively, their explanatory power was strongly dependent on the spatial scale with respect to both grain size and spatial extent. At the city-collective level (aggregating all studied cities), the adjusted $R^2$ of the full model with all three variables included demonstrated an increasingly linear-like (close to monotonic) scaling relation as the grain size coarsened from 0.25 km to 5 km (Figure 3a), and a decreasingly linear-like scaling relation as the spatial extent was enlarged from 20 km to 60 km (Figure 3e).

**Figure 3.** The power (adjusted $R^2$) of NDVI, water cover (WC), and POI density explaining the LST patterns across the grain size range from 0.25 km to 5 km (at each fixed spatial extent, a–d) and the spatial extent range from 20 to 60 km (at each fixed grain size, e–h).

The results from the individual variables revealed that their strengths in shaping the UHI patterns were also strongly scale-dependent at the city-collective level (Figure 3). With changing grain size, the strengths of the NDVI and POI density (measured by the adjusted $R^2$ of the univariate models) showed linear-like scaling relations (Figure 3b,d), which was similar to that of the total strength of all three variables (Figure 3a). However, their scaling relations with respect to spatial extent tended to exhibit hump-shaped patterns, which were clearly different to the total strength (Figure 3e). In contrast with the NDVI and POI, water cover had weaker strength, and its scaling relations with respect to both grain size and spatial extent seemed to have similar trends with (though more erratic than) those of the total strength (Figure 3c,g).

Among the three individual variables, NDVI generally had the highest relative importance at all spatial scales (Figure 4a), suggesting that vegetation cover is a dominant determinant of UHI patterns. Water cover and POI density appeared to have comparable relative importance at the city-collective level. We also observed a clear scale dependency of the relative importance of the variables. With increasing grain size, the relative importance of POI density gradually increased and exceeded that of water cover at coarse grain sizes of 2 km and 5 km. Moreover, we observed a strong scale dependence of the relative importance of the three variables. Generally, the relative importance of all three variables had linear-like scaling relations with respect to grain size (Figure 4a–c): the relative importance of NDVI and water cover declined while that of the POI density was elevated with coarsening grain sizes. With respect to spatial extent, in general, the relative importance of NDVI and POI presented a hump-shaped scaling relation, while water cover showed an inverse humped scaling relation.
4.2. Scaling Relations and Variable Relative Importance at the Individual-City Level

While the above-mentioned city-collective-level analysis clearly revealed the general scaling patterns, it remains unclear whether or not such ‘aggregated’ patterns have resulted from consistent scaling behavior across different studied cities. We then closely scrutinized each individual city. Compared with the linear (monotonic) scaling relations that were mostly found at the city-collective level, we observed much more diverse scaling relations at the individual-city level. Specifically, we identified six types of scaling relations, referred to as (1) monotonically increasing, (2) monotonically decreasing, (3) humped, (4) inversely humped, (5) scale invariant, and (6) erratic (Figure 5). The monotonic increasing or decreasing scaling relations were often the most frequent types across the individually studied cities (Figures 5 and 6, detailed scaling relations for each city could be found from Figures A5–A18 in Appendix B), giving rise to the general scaling patterns at the collective-city level. However, very contrasting scaling relations can also frequently arise. For example, the strength of the NDVI had both monotonic increasing and humped scaling relations with respect to grain size at the spatial extent of 20 km (Figure 5b, green bars). Our results thus demonstrate that the strengths of the UHI determinants as well as their relative importance can vary from city to city, suggesting that the scaling relations are clearly system-specific.

To further illustrate the system-specificity of the relative importance, we plotted the number of cities that were found to have NDVI, water cover, or POI density as the most important determinant at a given spatial scale (Figure 7, see Table A2, Appendix B for detailed information). The NDVI was consistently identified as the dominant determinant in ca. 20 cities (around 60%) across all observational scales; water cover was identified as the dominant determinant in up to 12 cities (i.e., ~35% of the studied cities); and POI density as the dominant determinant in up to seven cities (i.e., ~20% of the studied cities). An interesting pattern is that the dominance of POI density tended to be more pronounced at coarser grain sizes and larger spatial extents. In brief, the natural-surface variables were the most important UHI determinates, but the non-surface variable can occasionally dominate the UHI patterns, especially at larger spatial scales.
Figure 5. The frequency (number of cities) distributions of six types of scaling relations with respect to grain size (at each fixed spatial extent, a–d) and spatial extent (at each fixed grain size, e–h) for the individual explanatory power of NDVI, water cover (WC), and POI density as well as their overall explanatory power. The number of cities (the frequency) that had a specific scaling relation type is annotated above the corresponding bar.

Figure 6. The frequency (number of cities) distributions of six types of scaling relations with respect to grain size (at each fixed spatial extent, a–c) and spatial extent (at each fixed grain size, d–f) for the relative importance of NDVI, water cover (WC), and POI density. The number of cities (the frequency) that had a specific scaling relation type is annotated above the corresponding bar.
Figure 7. The frequency (number of cities) distributions of NDVI, water cover (WC), or POI density as the dominant urban-heat-island determinant that had the highest relative importance at each combination of spatial extent and grain size.

5. Discussion

5.1. Scale Dependency and System Specificity of Urban Heat Island Determinants

Our study demonstrates that the strengths of the UHI determinants as well as their relative importance are strongly dependent on the observational scale and are clearly system-specific. This is not very surprising in the sense that scale dependency and system specificity are generally expected to occur, and have been found in many geographic, ecological, and social patterns [30,31,43]. However, the remarkably high plasticity of the UHI determinants (to spatial scale and studied system) revealed by our analyses was indeed unexpected. For example, when comparing different spatial scales, we found up to a threefold difference in the assessed strength of particular determinants for the same city (see figures in Appendix B); and this difference can be even larger when comparing different cities. Such plasticity may easily lead to contrasting conclusions if the observation is based on a single scale and a single city, as in many previous studies [15,28,31].

A general scaling relation we found was that the selected natural-surface and non-surface variables, collectively, could better explain the observed LST patterns at smaller spatial extents, but coarser grain sizes. These detected scaling relations might be underpinned by the spatial heterogeneity of LST. With increasing spatial extent, the thermal environment would generally present higher spatial heterogeneity (caused by more heterogeneous topographic conditions, background climatic conditions, etc.) associated with more coexisting determinants and/or shaping mechanisms of UHIs [7,44,45]. When the grain size becomes finer, the local-scale variation of thermal conditions will become stronger as a result of increasing heterogeneity of micro-environment in terms of the spatial configuration of the buildings [10,46] three-dimensional (3D) vegetation structure [47,48], and other factors. These
heterogeneities may not be well captured by our selected variables and/or single linear models, leading to declined explanatory power.

For the individual variables, we found that the explanatory power of NDVI increased monotonically as the grain size became coarser. This finding is in line with some previous studies [11,49,50], but contradictory to others [26,44,51,52]. For instance, Weng, Lu, and Schubring [51] and Liu and Weng [52] suggested that the detected relationship between LST and NDVI was strongest at an intermediate grain size of 120 m in Indianapolis City, USA. While we cannot be certain whether this contradiction results from the different data sources and analysis methods used, it is very likely that it is attributed to system-specific scaling relations, because humped scaling relations of the NDVI were also found in some Chinese big cities in our study (Figure 5b). It has been documented that water bodies have a cooling effect [18,35,53], however, the scaling relations of water cover are not really clear. This multi-scale analysis revealed that 20 km is the threshold spatial extent for understanding the contribution of water cover to the LST pattern (Figure 3g), which further deepens the understanding of the contribution of water cover to LST and its cooling effect. Furthermore, we demonstrated that POI density can contribute to the formation of LST patterns and that its contribution increases with coarsening grain size.

In addition, studies on the formation mechanisms of LST patterns have mostly focused on particular influencing factors, leaving their relative importance poorly understood thus far, especially at different scales [13,26,48]. Natural surfaces and human activities can largely explain the LST pattern that was also demonstrated in this study, while the relative importance between both of them is controversial, as we stated above (see also in the Introduction). Through this multi-scale analysis, we found that the relative importance of green vegetation to LST was the highest at all spatial scales for the city-collective level, and it was significantly higher than the water cover and human activity intensity. This finding provides new evidence that the natural surface contributes more to the LST pattern than human activity [3,29]. However, the relative importance of water cover and POI to the LST shows different results at different scales (Figure 3). In contrast to the trend of human activity on the LST, the relative importance of water cover decreased with increasing grain size, and the spatial extent of 20 km is a threshold because the relative importance of water cover decreased significantly when the spatial extent exceeded 20 km (Figure 3). These results challenge our past understanding that the contribution of water cover to the LST generally increases with spatial scale, while the contribution of human activities to the LST is usually greater on a smaller spatial scale [6,26,35,53].

5.2. Implications

Many previous studies have demonstrated that green vegetation has a significant cooling effect, and the cooling effect can vary depending on the type of green vegetation, spatial pattern (composition and configuration) of green vegetation [12,17,18,49], and local background climate [54]. Our findings revealed that vegetation cover generally has the largest strength in shaping the LST patterns at most of the studied scales and lends support to previous argument that the natural surface contributes more to the LST pattern than human activity, and highlight the importance of green vegetation (and nature-based solution) for city planning in the context of climate adaptation [12]. In contrast, some case studies have found that the cooling effect of the water body was better than the green vegetation due to the higher rate of the heat consuming process of transpiration [14,54]. Our results suggest that this is also possible, as for the single-city level, the results of many cities showed that the water body was the dominant factor to explain the LST pattern, especially at finer spatial scales (Figure 6). This further indicated that the explanatory power and relative importance of the selected variables are largely scale-dependent and system-specific. Nevertheless, the results of the water body (Figures 2g and 3e) indicate that decision makers or planners can design water bodies on a smaller spatial grain and extent (20 km) to mitigate the UHI effect.

The study also showed that as the grain size and spatial extent increases, human activities can contribute to the LST pattern more (Figures 3 and 6). This result suggests that reducing anthropogenic
heat emissions is an optional measurement for climate mitigation. In particular, the urban agglomeration
has become the most salient feature of global urbanization in recent decades, hence regional thermal
environment mitigation has increasingly become the focus of policy attention. In addition, the results of
this study also showed that compared with the cooling effect of green space and water body, the relative
importance of human activities to LST was not that significant, especially at finer spatial scales (Figures 3
and 6). Therefore, according to these findings, the implication could be that compared with restricting
human activities induced anthropogenic heat emissions, planning more green vegetation (and water
bodies) is more efficient for climate adaption.

Overall, this study demonstrates that natural surfaces and human activities can largely determine
the LST pattern and that human activity has a positive effect on the increase in LST (SUHI). On the
other hand, to mitigate the UHI effect, we suggest that green vegetation (or nature-based solutions) is
the best way to cool cities and urban agglomerations. Moreover, from a larger perspective, we would
like to mention again the classic problem of landscape ecology–scale effect, that is, the patterns and
process relationships found in one system or one scale may not be applicable to another system or
scale. This study once again makes us notice the role of scale effects in other disciplines.

5.3. Limitations and Further Studies

Some limitations need to be mentioned in this study. First, due to image resolution limitations,
only five grain sizes from 0.25 km to 0.5 km were selected for this study. The city is a region with a
very high degree of landscape heterogeneity, so a high resolution of grain size is needed to obtain
more detailed information in follow up studies. At the same time, urban agglomeration has become
the most salient feature of urbanization in recent decades, so larger spatial scale-based studies are
also needed to provide valuable implications for climate adaption in urban agglomerations [8,16,55].
Second, previous studies have found that LST patterns vary between day and night and vary in
different seasons and climate zones. This study only used daytime thermal images to undertake
the study and did not take the background climate into account. Although the 32 selected Chinese
cities had similar meteorological conditions (e.g., temperature and relative humidity) in the summer
season, it would be better if we could take the background climate into consideration in future studies.
Therefore, future research should also further analyze the daytime and nighttime differences and
seasonal differences. Third, more detailed human activity data and the 3D structure of the city need to
be further considered. With the rapid development of information technology, more and more human
activity data have appeared, so we should try to use these more detailed and new data for further
research. Furthermore, urban 3D morphology also has effects on the LST pattern, but this study did
not consider the effects of urban 3D structure, which may underestimate the impact of human activities
on land surface temperature, particularly on air temperature.

In addition, although this study revealed the scale effect of the contribution and relative importance
of natural surfaces and human activities to the LST pattern as well as suggest implications for climate
adaption planning, the study still does not provide more valuable information such as how much
green vegetation and water should be placed on a given spatial scale to most effectively reduce UHIs.
Therefore, more research is needed to find the thresholds for the composition and configuration of
green vegetation and water cover on a given spatial scale.

6. Conclusions

Taking 32 selected Chinese cities as the sample, this study examined the scale (spatial extent and
grain size) effects of the contribution and relative importance of natural surfaces and human activities
to the UHI pattern. The study revealed that the contribution and relative importance of natural surfaces
and human activities are highly dependent on the scale-dependent and system-specific strengths.
Specifically, this study found the following. (1) The full model of natural surfaces and human activities
had a higher contribution to LST at all spatial scales generally, and the explanatory power and relative
importance of natural surfaces and human activities are scale-dependent and system-specific. (2) The
green space is negatively correlated with the LST pattern and showed more dominance than water and human activity intensity at all spatial scales, which implies that green vegetation (or nature-based solutions) is the best way to cool cities and urban agglomerations. (3) The water cover is significantly negatively correlated with the LST and the relative importance of water cover decreases with increasing grain size, and the spatial extent of 20 km is a threshold. Contrary to the water body, human activities are significantly positively correlated with the LST and as the spatial extent and grain size increase, human activities can contribute to the LST pattern more. These results can guide planners to plan water bodies on a smaller spatial grain and extent (20 km) to mitigate the UHI effect. Additionally, when conducting climate adaptation planning on a larger spatial scale, we should pay attention to reducing anthropogenic heat emissions, especially at a regional (urban agglomeration) scale. The study expands the understanding of the LST formation mechanisms at a different spatial scale and different cities, and its methods and results provide useful insights into thermal environment research and practice.

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Conflicts of Interest: The authors declare no conflict of interest.

Appendix A

Figure A1. An illustration of scaling analysis composed by changing spatial extent (a) and changing grain size (b).
Table A1. List of the cities and the path/row numbers and the time of the Landsat 8 images each city used.

| City       | Path/Row | Time       | City       | Path/Row | Time       |
|------------|----------|------------|------------|----------|------------|
| Anshan     | 119/31   | 2017.08.31 | Panjin     | 120/31   | 2017.09.07 |
| Baoding    | 123/33; 124/33 | 2017.07.10; 2017.07.01 | Shanghai 118/38; 118/39 | 2017.08.24; 2017.08.24 |
| Beijing    | 123/32   | 2017.07.10 | Shenyang   | 119/31   | 2017.08.31 |
| Changzhou  | 119/38   | 2017.05.27 | Shijiazhuang 124/33; 124/34 | 2016.08.31; 2016.08.31 |
| Datong     | 125/32   | 2017.08.25 | Suzhou     | 119/38   | 2017.05.27 |
| Fuxin      | 120/31   | 2017.09.07 | Tangshan   | 122/32; 122/33 | 2017.06.01; 2017.06.01 |
| Harbin     | 118/28   | 2017.07.07 | Tianjin    | 122/33   | 2015.06.12 |
| Handan     | 124/34; 124/35 | 2016.08.31; 2016.08.31 | Weifang 121/34; 121/35 | 2017.07.12; 2017.07.12 |
| Hefei      | 121/38   | 2016.07.25 | Wuhan      | 123/39   | 2016.07.23 |
| Hohhot     | 126/32   | 2017.06.29 | Xi’an      | 127/36   | 2016.06.17 |
| Jinan      | 122/34; 122/35 | 2015.06.12; 2015.06.12 | Xuzhou 121/36; 122/36 | 2015.06.05; 2015.07.30 |
| Linyi      | 121/36   | 2017.05.25 | Yinchuan   | 129/33   | 2016.07.01 |
| Luoyang    | 125/36   | 2017.08.09 | Changchun  118/29; 118/30 | 2016.07.04; 2016.07.04 |
| Nanchang   | 121/40   | 2017.09.14 | Changsha   123/40; 123/41 | 2016.07.23; 2016.07.23 |
| Nanjing    | 120/38   | 2017.07.21 | Zhengzhou  124/36 | 2015.09.14 |
| Nanning    | 125/44   | 2016.10.09 | Zibo       | 121/34; 121/35 | 2017.07.12; 2017.07.12 |
Figure A2. Map of the land surface temperature (LST) data for the 32 Chinese cities.

Figure A2. Cont.
Figure A2. Map of the land surface temperature (LST) data for the 32 Chinese cities.

Figure A2. Cont.
Figure A2. Map of the land surface temperature (LST) data for the 32 Chinese cities.
Figure A3. Map of the normalized difference vegetation index (NDVI) data and water cover (WC) data for the 32 Chinese cities.

Figure A3. Cont.
Figure A3. Map of the normalized difference vegetation index (NDVI) data and water cover (WC) data for the 32 Chinese cities.
Figure A3. Map of the normalized difference vegetation index (NDVI) data and water cover (WC) data for the 32 Chinese cities.
Figure A4. Map of the point of interest (POI) data for the 32 Chinese cities.

Figure A4. Cont.
Figure A4. Map of the point of interest (POI) data for the 32 Chinese cities.

Figure A4. Cont.
Figure A4. Map of the point of interest (POI) data for the 32 Chinese cities.

Appendix B

Figures of the scalograms of the explanatory power (adjusted $R^2$) for each variable (NDVI, WC, POI) and their collective explanatory power of linear models at different grain size (Figures A5–A8) and different spatial extent (Figures A9–A12); Figures of scalograms of the relative importance for each variable (NDVI, WC, POI) at different grain size (Figures A13–A15) and different spatial extent (Figures A16–A18)
Figure A5. Explanatory power of NDVI for explaining LST (Scalogram of the grain size).
Figure A6. Explanatory power of WC for explaining LST (Scalogram of the grain size).
Figure A7. Explanatory power of POI for explaining LST (Scalogram of the grain size).
Figure A8. Collective explanatory power for explaining LST (Scalogram of the grain size).
Figure A9. Explanatory power of NDVI for explaining LST (Scalogram of the spatial extent).
Figure A10. Explanatory power of WC for explaining LST (Scalogram of the spatial extent).
Figure A11. Explanatory power of POI for explaining LST (Scalogram of the spatial extent).
Figure A12. Collective explanatory power for explaining LST (Scalogram of the spatial extent).
Figure A13. Relative importance of NDVI for explaining LST (Scalogram of the grain size).
Figure A14. Relative importance of WC for explaining LST (Scalogram of the grain size).
Figure A15. Relative importance of POI for explaining LST (Scalogram of the grain size).
Figure A16. Relative importance of NDVI for explaining LST (Scalogram of the spatial extent).
Figure A17. Relative importance of WC for explaining LST (Scalogram of the spatial extent).
Figure A18. Relative importance of NDVI for explaining LST (Scalogram of the spatial extent).
Table A2. Count of the cities for each factor that should the most importance for explaining the LST pattern under different scales.

|       | 0.25 km |       |       |       |       |       |       |       |       |       |       |       |       |       |       |       |
|-------|---------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|
|       | NDVI    | WC    | POI   | NDVI  | WC    | POI   | NDVI  | WC    | POI   | NDVI  | WC    | POI   | NDVI  | WC    | POI   |
| 20 km | 21      | 11    | 0     | 20    | 12    | 0     | 19    | 10    | 3     | 19    | 10    | 3     | 19    | 9     | 4     |
| 30 km | 21      | 11    | 0     | 22    | 10    | 0     | 23    | 8     | 1     | 20    | 7     | 5     | 21    | 7     | 4     |
| 40 km | 23      | 9     | 0     | 23    | 8     | 1     | 21    | 7     | 4     | 20    | 7     | 5     | 23    | 4     | 5     |
| 50 km | 23      | 9     | 0     | 23    | 8     | 1     | 22    | 5     | 5     | 21    | 4     | 7     | 23    | 4     | 5     |
| 60 km | 24      | 8     | 0     | 23    | 6     | 3     | 22    | 4     | 6     | 22    | 4     | 6     | 21    | 4     | 7     |
Figure A19. The power (pseudo $R^2$, the non-spatial component of simultaneous autoregressive) of NDVI, WC and POI density explaining the LST patterns across the grain size from 0.25 km to 5 km (at each fixed spatial extent, upper panels) and the spatial extent range from 20 to 60 km (at each fixed grain size, lower panels).
References

1. United Nations. World Urbanization Prospects: The 2018 Revision (st/esa/ser.a/420); United Nations, Department of Economic and Social Affairs, Population Division: New York, NY, USA, 2019.
2. Oke, T.R. Boundary Layer Climates; Routledge: London, UK; New York, NY, USA, 2002.
3. Forman, R.T. Urban Ecology: Science of Cities; Cambridge University Press: Cambridge, UK, 2014.
4. Akbari, H.; Kolokotsa, D. Three decades of urban heat islands and mitigation technologies research. Energy Build. 2016, 133, 834–842. [CrossRef]
5. Foley, J.A.; DeFries, R.; Asner, G.P.; Barford, C.; Bonan, G.; Carpenter, S.R.; Chapin, F.S.; Coe, M.T.; Daily, G.C.; Gibbs, H.K. Global consequences of land use. Science 2005, 309, 570–574. [CrossRef] [PubMed]
6. Sun, R.; Wang, Y.; Chen, L. A distributed model for quantifying temporal-spatial patterns of anthropogenic heat based on energy consumption. J. Clean. Prod. 2018, 170, 601–609. [CrossRef]
7. Buyantuyev, A.; Wu, J. Urban heat islands and landscape heterogeneity: Linking spatiotemporal variations in surface temperatures to land-cover and socioeconomic patterns. Landsc. Ecol. 2010, 25, 17–33. [CrossRef]
8. Yu, Z.; Yao, Y.; Yang, G.; Wang, X.; Vejre, H. Spatiotemporal patterns and characteristics of remotely sensed regional heat islands during the rapid urbanization (1995–2015) of southern China. Sci. Total Environ. 2019, 674, 242–254. [CrossRef] [PubMed]
9. Churkina, G.; Kuik, F.; Bonn, B.; Lauer, A.; Grote, R.; Tomiak, K.; Butler, T.M. Effect of VOC emissions from vegetation on air quality in Berlin during a heatwave. Environ. Sci. Technol. 2017, 51, 6120–6130. [CrossRef]
10. Huang, G.; Cadenasso, M.L. People, landscape, and urban heat island: Dynamics among neighborhood social conditions, land cover and surface temperatures. Landsc. Ecol. 2016, 31, 2507–2515. [CrossRef]
11. Estoque, R.C.; Murayama, Y.; Myint, S.W. Effects of landscape composition and pattern on land surface temperature: An urban heat island study in the megacities of Southeast Asia. Sci. Total Environ. 2017, 577, 349–359. [CrossRef]
12. Yu, Z.; Guo, X.; Jørgensen, G.; Vejre, H. How can urban green spaces be planned for climate adaptation in subtropical cities? Ecol. Indic. 2017, 82, 152–162. [CrossRef]
13. Peng, J.; Xie, P.; Liu, Y.; Ma, J. Urban thermal environment dynamics and associated landscape pattern factors: A case study in the Beijing metropolitan region. Remote Sens. Environ. 2016, 173, 145–155. [CrossRef]
14. Santamouris, M.; Ban-Weiss, G.; Osmund, P.; Paolini, R.; Synnefa, A.; Cartalis, C.; Muscio, A.; Zinzi, M.; Morakinyo, T.E.; Ng, E.; et al. Progress in urban greenery mitigation science—assessment methodologies advanced technologies and impact on cities. J. Civ. Eng. Manag. 2018, 24, 638–671. [CrossRef]
15. Li, X.; Zhou, W.; Ouyang, Z. Relationship between land surface temperature and spatial pattern of greenspace: What are the effects of spatial resolution? Landsc. Urban Plan. 2013, 114, 1–8. [CrossRef]
16. Yu, Z.; Yao, Y.; Yang, G.; Wang, X.; Vejre, H. Strong contribution of rapid urbanization and urban agglomeration development to regional thermal environment dynamics and evolution. For. Ecol. Manag. 2019, 446, 214–225. [CrossRef]
17. Fan, H.; Yu, Z.; Yang, G.; Liu, T.Y.; Liu, T.Y.; Hung, C.H.; Vejre, H. How to cool hot-humid (Asian) cities with urban trees? An optimal landscape size perspective. Agric. For. Meteorol. 2019, 265, 338–348. [CrossRef]
18. Yang, G.; Yu, Z.; Jørgensen, G.; Vejre, H. How can urban blue-green space be planned for climate adaption in high-latitude cities? A seasonal perspective. Sustain. Cities Soc. 2020, 53, 101932. [CrossRef]
19. Kuang, W.; Liu, Y.; Dou, Y.; Chi, W.; Chen, G.; Gao, C.; Yang, T.; Liu, J.; Zhang, R. What are hot and what are not in an urban landscape: Quantifying and explaining the land surface temperature pattern in Beijing, China. Landsc. Ecol. 2015, 30, 357–373. [CrossRef]
20. Xu, H. Modification of normalised difference water index (ndwi) to enhance open water features in remotely sensed imagery. Int. J. Remote Sens. 2006, 27, 3025–3033. [CrossRef]
21. Yu, Z.; Guo, X.; Zeng, Y.; Koga, M.; Vejre, H. Variations in land surface temperature and cooling efficiency of green space in rapid urbanization: The case of Fuzhou City, China. Urban For. Urban Green. 2018, 29, 113–121. [CrossRef]
22. Liu, Y.; Liu, X.; Gao, S.; Gong, L.; Kang, C.; Zhi, Y.; Chi, G.; Shi, L. Social sensing: A new approach to understanding our socioeconomic environments. Ann. Assoc. Am. Geogr. 2015, 105, 512–530. [CrossRef]
23. Raghupathi, W.; Raghupathi, V. Big data analytics in healthcare: Promise and potential. Health Inf. Sci. Syst. 2014, 2, 3. [CrossRef]
24. Wei, S.; Teng, S.N.; Li, H.J.; Xu, J.; Ma, H.; Luan, X.L.; Yang, X.; Shen, D.; Liu, M.; Huang, Z.Y.; et al. Hierarchical structure in the world’s largest high-speed rail network. *PLoS ONE* **2019**, *14*, e0211052. [CrossRef] [PubMed]

25. McKenzie, G.; Janowicz, K.; Adams, B. A weighted multi-attribute method for matching user-generated points of interest. *Cartogr. Geogr. Inf. Sci.* **2014**, *41*, 125–137. [CrossRef]

26. Luan, X.; Wei, S.; Han, S.; Li, X.; Yang, W.; Liu, M.; Xu, C. A multi-scale study on the formation mechanism and main controlling factors of urban thermal field based on urban big data. *Chin. J. Appl. Ecol.* **2018**, *29*, 2861–2868.

27. Sun, Y.; Gao, C.; Li, J.; Wang, R.; Liu, J. Quantifying the effects of urban form on land surface temperature in subtropical high-density urban areas using machine learning. *Remote Sens.* **2019**, *11*, 959. [CrossRef]

28. Ma, Q.; Wu, J.; He, C. A hierarchical analysis of the relationship between urban impervious surfaces and land surface temperatures: Spatial scale dependence, temporal variations, and bioclimatic modulation. *Landsc. Ecol.* **2016**, *31*, 1–15. [CrossRef]

29. Stewart, I.D.; Oke, T.R. Local climate zones for urban temperature studies. *Bull. Am. Meteorol. Soc.* **2012**, *93*, 1879–1900. [CrossRef]

30. Du, H.; Wang, D.; Wang, Y.; Zhao, X.; Qin, F.; Jiang, H.; Cai, Y. Influences of land cover types, meteorological conditions, anthropogenic heat and urban area on surface urban heat island in the yangtze river delta urban agglomeration. *Sci. Total Environ.* **2016**, *571*, 461–470. [CrossRef]

31. Wu, J. Effects of changing scale on landscape pattern analysis: Scaling relations. *Landsc. Ecol.* **2004**, *19*, 125–138. [CrossRef]

32. ESA Climate Change Initiative. Land Cover cci, Product User Guide, Version 2.0. Cci-lc-Pugv2. Available online: [http://maps.Ellie.Ucl.Ac.Be/cci/viewer](http://maps.Ellie.Ucl.Ac.Be/cci/viewer) (accessed on 16 December 2017).

33. Jiménez-Muñoz, J.C.; Sobrino, J.A.; Skoković, D.; Mattar, C.; Cristóbal, J. Land surface temperature retrieval methods from landsat-8 thermal infrared sensor data. *IEEE Geosci. Remote Sens. Lett.* **2014**, *11*, 1840–1843. [CrossRef]

34. Atmospheric Correction Parameter Calculator. Available online: [http://atmcorr.gsfc.nasa.gov/](http://atmcorr.gsfc.nasa.gov/) (accessed on 31 December 2017).

35. Ellison, D.; Morris, C.E.; Locatelli, B.; Sheil, D.; Cohen, J.; Murdiyarso, D.; Gutierrez, V.; Noordwijk, M.V.; Creed, I.F.; Pokorny, J.; et al. Trees, forests and water: Cool insights for a hot world. *Glob. Environ. Chang.* **2017**, *43*, 51–61. [CrossRef]

36. AMAP. Available online: [https://ditu.amap.com/](https://ditu.amap.com/) (accessed on 31 October 2017).

37. Yuan, F.; Bauer, M.E. Comparison of impervious surface area and normalized difference vegetation index as indicators of surface urban heat island effects in landsat imagery. *Remote Sens. Environ.* **2007**, *106*, 375–386. [CrossRef]

38. Tonidandel, S.; LeBreton, J.M. Relative importance analysis: A useful supplement to regression analysis. *J. Bus. Psychol.* **2011**, *26*, 1–9. [CrossRef]

39. Kissling, W.D.; Carl, G. Spatial autocorrelation and the selection of simultaneous autoregressive models. *Glob. Ecol. Biogeogr.* **2008**, *17*, 59–71. [CrossRef]

40. Teng, S.N.; Xu, C.; Sandel, B.; Svenning, J.C. Effects of intrinsic sources of spatial autocorrelation on spatial regression modelling. *Methods Ecol. Evol.* **2018**, *9*, 363–372. [CrossRef]

41. Xu, C.; Huang, Z.Y.X.; Chi, T.; Chen, B.J.W.; Zhang, M.; Liu, M. Can local landscape attributes explain species richness patterns at macroecological scales? *Glob. Ecol. Biogeogr.* **2014**, *23*, 436–445. [CrossRef]

42. Tonidandel, S.; Lebreton, J.M. Rwa web: A free, comprehensive, web-based, and user-friendly tool for relative weight analyses. *J. Bus. Psychol.* **2015**, *30*, 207–216. [CrossRef]

43. Allen, T.F.; Starr, T.B. *Hierarchy: Perspectives for Ecological Complexity*; University of Chicago Press: Chicago, IL, USA, 2017.

44. Sun, R.; Lü, Y.; Yang, X.; Chen, L. Understanding the variability of urban heat islands from local background climate and urbanization. *J. Clean. Prod.* **2019**, *208*, 743–752. [CrossRef]

45. Zhao, L.; Lee, X.; Smith, R.B.; Oleson, K. Strong contributions of local background climate to urban heat islands. *Nature* **2014**, *511*, 216–219. [CrossRef]

46. Yin, C.; Yuan, M.; Lu, Y.; Huang, Y.; Liu, Y. Effects of urban form on the urban heat island effect based on spatial regression model. *Sci. Total Environ.* **2018**, *634*, 696–704. [CrossRef]
47. Zhang, Y.; Middel, A.; Turner, B.L. Evaluating the effect of 3d urban form on neighborhood land surface temperature using google street view and geographically weighted regression. *Lands. Ecol.* 2019, 34, 681–697. [CrossRef]

48. Shiflett, S.A.; Liang, L.L.; Crum, S.M.; Feyisa, G.L.; Wang, J.; Jenerette, G.D. Variation in the urban vegetation, surface temperature, air temperature nexus. *Sci. Total Environ.* 2019, 34, 681–697. [CrossRef] [PubMed]

49. Zhou, W.; Wang, J.; Cadenasso, M.L. Effects of the spatial configuration of trees on urban heat mitigation: A comparative study. *Remote Sens. Environ.* 2017, 195, 1–12. [CrossRef]

50. Han, S.R.; Wei, S.; Zhou, W.; Zhang, M.J.; Tao, T.T.; Qiu, L.; Liu, M.S.; Xu, C. Quantifying the spatial pattern of urban thermal fields based on point of interest data and landsat images. *Acta Ecol. Sin.* 2017, 37, 5305–5312.

51. Weng, Q.; Lu, D.; Schubring, J. Estimation of land surface temperature–vegetation abundance relationship for urban heat island studies. *Remote Sens. Environ.* 2004, 89, 467–483. [CrossRef]

52. Liu, H.; Weng, Q. Scaling effect on the relationship between landscape pattern and land surface temperature. *Photogramm. Eng. Remote Sens.* 2009, 75, 291–304. [CrossRef]

53. Du, H.; Song, X.; Jiang, H.; Kan, Z.; Wang, Z.; Cai, Y. Research on the cooling island effects of water body: A case study of shanghai, china. *Ecol. Indic.* 2016, 67, 31–38. [CrossRef]

54. Yu, Z.; Xu, S.; Zhang, Y.; Jørgensen, G.; Vejre, H. Strong contributions of local background climate to the cooling effect of urban green vegetation. *Sci. Rep.* 2018, 8, 6798. [CrossRef]

55. Gao, J.; Yu, Z.; Wang, L.; Vejre, H. Suitability of regional development based on ecosystem service benefits and losses: A case study of the yangtze river delta urban agglomeration, china. *Ecol. Indic.* 2019, 107, 105579. [CrossRef]

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