Multi-scale Assessment of Agricultural Consumptive Water Use in California’s Central Valley

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Key Points:
- A 30 m daily evapotranspiration product estimated using regionally optimized approach and Landsat Analysis Ready Data Collection 1.
- Assessment of agricultural consumptive water use over Central Valley, provides critical guidance for sustainable groundwater management.
- Large variation of water use was found, mostly due to crop diversity, age structure, and physiological factors.
Abstract
Spatial estimates of crop evapotranspiration with high accuracy from the field to watershed scale have become increasingly important for water management, particularly over irrigated agriculture in semi-arid regions. Here, we provide a comprehensive assessment on patterns of annual agricultural water use over California’s Central Valley, using 30-m daily evapotranspiration estimates based on Landsat satellite data. A semi-empirical Priestley-Taylor approach was locally optimized and cross-validated with available field measurements for major crops including alfalfa, almond, citrus, corn, pasture, and rice. The evapotranspiration estimates explained more than 70% variance in daily measurements from independent sites with an RMSE of 0.88 mm day\(^{-1}\). When aggregated over the Valley, we estimated an average evapotranspiration of 820 ± 290 mm yr\(^{-1}\) in 2014. Agricultural water use varied significantly across and within crop types, with a coefficient of variation ranging from 8% for Rice (1110 ± 85 mm yr\(^{-1}\)) to 59% for Pistachio (592 ± 352 mm yr\(^{-1}\)). Total water use in 2016 increased by 9.6%, as compared to 2014, mostly because of land-use conversion from fallow/idle land to cropland. Analysis across 134 Groundwater Sustainability Agencies (GSAs) further showed a large variation of agricultural evapotranspiration among and within GSAs, especially for tree crops, e.g., almond evapotranspiration ranging from 339 ± 80 mm yr\(^{-1}\) in Tracy to 1240 ± 136 mm yr\(^{-1}\) in Tri-County Water Authority. Continuous monitoring and assessment of the dynamics and spatial heterogeneity of agricultural evapotranspiration provide data-driven guidance for more effective land use and water planning across scales.

1 Introduction
Evapotranspiration, also known as crop consumptive water use in agricultural lands, is often the largest yet most uncertain component of the agricultural water balance in semi-arid regions (Burt et al., 2002; Fisher et al., 2017; Oki et al., 2006). Irrigation accounts for 70% of human freshwater use and supports 45% of the world’s food supply production (Koech & Langat, 2018). Most irrigation water is lost to the atmosphere via evapotranspiration (California Department of Water Resources, 2020a). Information about evapotranspiration can be applied to optimized sustainable regional water allocations and farm-level irrigation management, improving water and food security in a changing climate to meet the demands of a growing population (Fisher et al., 2017).

As one of the world’s most productive agricultural regions, California is the major producer of various specialty crops including half of the fruits, vegetables, and nuts grown in the United States (California Department of Food and Agriculture, 2018). A quarter of the nation’s produce comes from the Central Valley alone, due to its fertile soils and extended growing season, generating $20 billion of agricultural sales in 2017 (U.S. Department of Agriculture - National Agricultural Statistics Service, 2019). More than 7 million acres (28,000 square kilometers) of the Central Valley are irrigated via a system of reservoirs and canals, accounting for about 75% of the state’s irrigated lands and consuming about 23 trillion liters of water per year (Faunt, 2009). The acreage of high-value yet water-intensive tree crops, such as almonds, pistachios, and walnuts, has been expanding in recent decades, increasing water demand in the state’s largest groundwater basin (Strom, 2014). Groundwater is a major source of water supply in the Central Valley, ranging from 30% during wet years to 70% during extremely dry years (Faunt et al., 2016). During the recent 2012-2016 drought, groundwater pumping was intensified, which led to the doubling of the reduction in the valley’s groundwater storage to 11 trillion liters.
per year from the 2007-2009 value (Xiao et al., 2017). This groundwater overdraft due to
excessive pumping accounts for 13% of water sources in the San Joaquin Valley, which is the
southern half of the Central Valley (Hanak et al., 2017). It caused significant land subsidence
(Faunt et al., 2016), and subsequent damage on aqueducts, costing the state “tens of millions of
dollars” in repairs to the aqueduct in the last 40 years (Smith, 2015).

The intensifying drought and reduction in water supply during the summer-growing-
season are expected to continue, due to a projected decrease in snowpack and increasing
temperature in the coming decades (Cvijanovic et al., 2017; Lynn, 2015). Consequences of
chronic groundwater overdraft and non-point sources pollution motivated the enactment of the
Sustainable Groundwater Management Act in 2014. It requires local groundwater agencies in
critically over-drafted basins to achieve a sustainable water balance by 2040 (California
Department of Water Resources, 2019c). Another management challenge under water scarcity is
to balance the competing beneficial water uses by many stakeholders, e.g., maintaining river
flows to support estuarine habitat (Escriva-Bou et al., 2019). Some potential water management
alternatives to attain Sustainable Groundwater Management Act requirements include demand
management, e.g., via reduced irrigated area, in addition to increasing supply via groundwater
recharge. Spatial evapotranspiration estimate records could inform stakeholders on historical and
present water demand. Time series of accurate evapotranspiration maps can also help water
managers better understand the dynamics of groundwater pumping, improve regulation
oversight, and evaluate the impacts of the implemented water policy (Garcia et al., 2016; Hanak,
2011). As 77% of cropland owners in California have fields smaller than 0.4 km² (Macaulay &
Butsic, 2017), remotely monitoring individual land owner’s water use is only possible by using
routine satellite observations at a relatively high spatial resolution.

A few algorithms of varying complexity have been developed to map crop

evapotranspiration using remote sensing data, due to the large spatial coverage and consistent
imagery acquisition by satellite instruments (Akbar et al., 2019; Allen et al., 2007; Anderson et
al., 2012; Bastiaanssen et al., 1998; Melton et al., 2012; Nagler et al., 2005; Norman et al., 1995)
(Table S1). For example, the crop coefficient-based evapotranspiration approach has been used
by California’s growers for irrigation management (California Department of Water Resources,
2005; Melton et al., 2012). This method is relatively easy to implement but does not account for
other factors, such as water stress and thus likely overestimates actual evapotranspiration
(Anderson et al., 2012; Cuenca et al., 2013). Energy balance-based approaches, on the other
hand, estimate evapotranspiration as a residual between available energy and sensible heat flux
(Allen et al., 2007; Bastiaanssen et al., 1998; Norman et al., 1995; Su, 2002), e.g., as adopted by
the Surface Energy Balance Algorithm for Land (SEBAL) (Bastiaanssen et al., 1998) and
Mapping EvapoTranspiration at high Resolution with Internalized Calibration (METRIC) (Allen
et al., 2007). In METRIC, the sensible heat is often estimated using clusters of hot and cold
pixels observed within each Landsat scene (Allen et al., 2007). The pixel selection can be fully
automated (Allen et al., 2013) for large-scale applications or manually selected by professional
users for more accurate regional estimates.

More sophisticated methods involve solving the soil and canopy latent heat components
within a process model, constrained by satellite data (Anderson et al., 1997, 2018; Ryu et al.,
2011). The Atmosphere-Land EXchange Inverse (ALEXI) model, for example, implements the
Two-Source Energy Balance in a time differential mode, based on two snapshots of high
temporal frequency geostationary satellite thermal observations, to reduce the sensitivity to the
errors in absolute land surface temperatures (Anderson et al., 1997, 2004, 2018). DisALEXI has
been further developed to estimate evapotranspiration at 30 m resolution, by bringing in additional higher spatial resolution thermal data from Landsat (Anderson et al., 2004, 2012; Knipper et al., 2018; Norman et al., 2003). A recent comparative study of crop evapotranspiration estimates over the Sacramento San Joaquin Delta showed that DisALEXI estimates had an RMSE of 1.43 mm day\(^{-1}\) and mean bias of 0.13 mm day\(^{-1}\) while the calibrated METRIC prepared by the Irrigation Training & Research Center had an RMSE and mean bias of 2.55 and 2.06 mm day\(^{-1}\), respectively (Medellín-Azuara, J. et al., 2018). Another biophysical model, Breathing Earth System Simulator (BESS) (Baldocchi et al., 2019; Jiang & Ryu, 2016; Ryu et al., 2011), couple surface energy balance, photosynthesis, and stomatal conductance processes, to estimate evapotranspiration, forced by biophysical parameters from MODIS 8-day 1-km observations.

One compromise between the simple crop-coefficient based approaches and more complex energy balance models described above is the Priestley Taylor approach. It estimates evapotranspiration over the wet and extensive surface (Debruin H. A. R., 1983; Eichinger et al., 1996; Priestley & Taylor, 1972). It is primarily driven by available energy and a Priestley Taylor coefficient that partitions available energy to latent and sensible heat. This coefficient was found to be constant for estimating potential evapotranspiration. For drying environmental conditions, Priestley & Taylor (1972) demonstrated that a factor could be added to multiply the Priestley Taylor evapotranspiration estimates to actual evapotranspiration. They found that the factor remains constant when the cumulative evapotranspiration over a drying surface was below a certain threshold, but the threshold varies from one site to another. After the threshold was exceeded, the scaling factor follows a linear decline to zero after a further 5cm of evapotranspiration. Similar to Priestley & Taylor (1972)’s concept of scaling Priestley Taylor evapotranspiration estimates on wet surfaces to actual evapotranspiration over drying surfaces, remote sensing scientists introduced biophysical controls, such as water stress, to down-regulate the Priestley Taylor coefficient, thus reducing the fraction of available energy used for latent heat. Fisher et al. (2008), for example, parameterized the Priestley Taylor coefficient separately for soil evaporation, canopy transpiration, and interception using remotely sensed vegetation index and meteorological data. It has been applied to estimate monthly evapotranspiration globally at 5 km and 1° resolutions using Advanced Very High Resolution Radiometer data (Fisher et al., 2008; Vinukollu et al., 2011), regionally at 1km using Moderate Resolution Imaging Spectroradiometer (MODIS) data (Yao et al., 2015) and most recently at 70 m using ECOSystem Spaceborne Thermal Radiometer Experiment on Space Station data (Fisher et al., 2020). Jin et al. (2011) optimized Priestley Taylor coefficient, \(\alpha\), as a function of Leaf Area Index (LAI) and soil moisture for each plant function type, using the eddy covariance tower measurements from AmeriFlux sites, and estimated monthly evapotranspiration at 1km for the entire continental US, from primarily MODIS data.

In this study, our primary objective is to improve our understanding of the agricultural water use patterns in California’s Central Valley, to facilitate water resources planning, in the context of the Sustainable Groundwater Management Act. We first modified, calibrated, and evaluated the refined semi-empirical Priestley Taylor method (Jin et al., 2011) over major California crops, using the Landsat Analysis Ready Data (Dwyer et al., 2018). This approach further built upon an earlier version adapted for evapotranspiration estimation in Sacramento San Joaquin Delta (Medellín-Azuara, J. et al., 2018). The method was chosen for this study based on its relatively good performance and computational efficiency for large-scale applications (Medellín-Azuara, J. et al., 2018). The automated workflow for regional applications was applied
to assess the spatial patterns of evapotranspiration across the Central Valley and among

dominant crop types and to analyze water use changes between 2014 and 2016 water years. To

provide insights for water management agencies in their efforts to prepare for droughts and

implement Sustainable Groundwater Management Act, we further provided a comprehensive

assessment on agricultural evapotranspiration by boundaries of Groundwater Sustainability

Agencies (GSAs).

2 Data and Methods

We adapted the semi-empirical Priestley-Taylor method (Jin et al., 2011), as shown in

Equation (1), for agricultural sites in California. The method is driven by Landsat remote sensing
data and was calibrated and evaluated using recently available in-situ evapotranspiration

measurements, from 2003 to 2019, over a variety of irrigated crop types.

\[ LE = PT_a \times \frac{\Delta}{\Delta+\gamma} \times (R_n - G) \]  

(1)

\( \Delta \) represents slope of vapor pressure – temperature curve and \( \gamma \) psychrometric constant. Net

radiation \( (R_n) \) and ground heat flux \( (G) \) were estimated from Landsat observations. \( PT_a \), hereafter

referred to as the actual Priestley Taylor coefficient, approximates the fraction of available

energy used for actual evapotranspiration. For potential evapotranspiration without any water

limitation, the potential Priestley Taylor coefficient was found to be a constant near 1.26 (Paw U

& Gao, 1988). For actual evapotranspiration, we here estimated \( PT_a \) as a function of plant growth

and water stress, as shown in section 2.4.

2.1 Study area

We focused on the major agricultural production area, about 25 thousand km\(^2\), in the

Central Valley of California (Figure 1a). The top 8 crop types include almond, grapes, corn, rice,

alfalfa, walnuts, pistachios, and tomatoes, based on the updated crop layers from the California

Department of Water Resources (DWR) (California Department of Water Resources, 2020b)

(Figures 9a, c, 10a). Together, these account for 68 and 73% of agricultural land use in 2014 and

2016, respectively. This semi-arid region has a Mediterranean climate, with mild winter and hot,
dry summers. Its mean annual precipitation ranges from 51 cm yr\(^{-1}\) in the north to 13 cm yr\(^{-1}\) in

the far south, and the majority of rainfall occurs from November to March. The Central Valley is

therefore highly dependent on irrigation and vulnerable to water scarcity. Its groundwater storage

depleted by approximately 16 trillion liters between spring 2005 and 2010 (California

Department of Water Resources, 2019a). Groundwater depletion during the recent 2012-2016

drought is expected to be even worse, since Xiao et al., (2017) estimated that the Central

Valley’s groundwater storage depletion doubled to 11 trillion liters per year from the 2007-2009

value.

GSAs were established under the Sustainable Groundwater Management Act to manage

California’s groundwater resources at the local scale. Over 260 GSAs were initially formed in

the State’s high and medium priority and some are currently under reorganization and

consolidation. Figure 1c shows 134 GSAs in the Central Valley that have more than 1 km\(^2\) of

farmland in both 2014 and 2016. The GSAs boundaries are defined by the most recently

available DWR’s Exclusive GSA map (California Department of Water Resources, 2019b);

GSAs boundaries were sometimes split into parts by other political boundaries, so we merged the

polygon parts that share an identical GSA name, e.g. Byron-Bethany Irrigation District GSA -
East Contra Costa and Byron-Bethany Irrigation District GSA – Tracy were merged into a single GSA.

2.2 Input data

2.2.1 Field measurements

A total of 26 evapotranspiration measurement sites over agricultural areas were available in the study region, including five agricultural sites from the AmeriFlux network and two from cropland sites established for the shorter-term study of California’s specialty crops (Table 1). The measurement sites represented 8 crop types, namely, alfalfa, almond, citrus, corn, pasture, rice, tomato, and beardless winter wheat (Table 1). Evapotranspiration was measured either with an eddy covariance system (Baldocchi et al., 1988; Paw U et al., 2000; Swinbank, 1951) or with a surface renewal system (Paw U et al., 1995). The eddy covariance system uses a sonic anemometer and infrared gas analyzer to measure three-dimensional wind velocities and high-frequency fluctuations of water vapor concentrations (Baldocchi et al., 2001). It measures evapotranspiration by monitoring the vertical flux of water vapor. High-frequency eddy covariance measurements in 2 alfalfa, 2 corn, and one rice AmeriFlux sites were collected and preprocessed into half-hourly evapotranspiration data as outlined in Eichelmann et al., 2018 and Hemes et al., 2019. Of the five AmeriFlux sites, net radiation for alfalfa and corn was measured with four-channel net radiometers.

Most sites were located in the Sacramento San Joaquin Delta region were also employed in the Delta Consumptive Use Comparative Study (Medellín-Azuara, J. et al., 2018) supported by the California State Water Resources Control Board Office of the Delta Watermaster and other agencies. Less expensive surface renewal systems were deployed over 14 sites for corn, alfalfa, and pasture. They use thermocouples to measure sensible heat flux, an NRLITE2 Net Radiometer for net radiation, and either measure ground heat flux \((G)\) with a combination of ground heat flux plates and soil thermocouples or assume it is zero for daily estimates. Evapotranspiration is then estimated as the residual of the energy balance. For each crop type, an eddy covariance tower was deployed to calibrate the sensible heat flux relationship between eddy covariance and surface renewal measurements (Castellví et al., 2006; Shapland et al., 2012).

Evapotranspiration measurements were compiled from two specialty crop research projects in Tulare and Kern county of the southern Central Valley, including surface renewal measurements in citrus orchards from 2001 to 2004 (Consoli et al., 2006; Snyder & O’Connell, 2007) and eddy covariance measurements in an almond orchard from 2009 to 2012 (Falk et al., 2013; He et al., 2017). We used only data collected after February 2003 in this study, considering the data availability of California Irrigation Management Information System Spatial product (Spatial-CIMIS) data. The most recently available eddy covariance tower measurements by NASA JPL were also added. The JPL sites were located at the Russell Ranch research field, near Davis, including one over tomato from February to October 2017, and the other over winter wheat from December 2016 to October 2017. These towers have advanced thermal infrared radiometers to measure land surface temperature, and two sets of four channels net radiometers to reduce measurement uncertainty. High-frequency evapotranspiration data were automatically processed using Campbell Scientific Inc.’s standard Eddy-Covariance Datalogger Program software and various quality control procedures. All half-hourly measurements were preprocessed and aggregated into daily evapotranspiration if less than 20% of the half-hourly measurements were missing within a day.
2.2.2 Meteorological data

We obtained the daily gridded meteorological data, including minimum and maximum-air temperature at 1.5 m, and daily dew point, from Spatial-CIMIS at a 2 km resolution (Hart et al., 2009). The DWR manages a network of over 145 automated weather stations over well-maintained and well-watered grass sites across California providing reference evapotranspiration for pasture. The station data were spatially interpolated to produce the 2-km gridded dataset since 2003. We also used the Spatial-CIMIS cloud cover and incoming solar radiation for both clear-sky and all-sky conditions, derived from Geostationary Operational Environmental Satellite visible channel imager data, for our radiation component calculation.

2.2.3 Remote Sensing data

All available surface reflectance and surface temperature products, and the corresponding quality assessment layers at 30 m were downloaded from USGS Landsat Analysis Ready Dataset (Dwyer et al., 2018). A total of 8 tiles covered the whole study area. The land surface temperature retrieval from the Landsat thermal data is based on a radiative transfer model with an improved surface emissivity estimate (Cook, 2014). Each active Landsat satellite takes snapshots between 9:53 to 10:55 am Pacific Standard Time every 16 days. Invalid or high uncertainty pixel values were filtered based on the quality assessment rasters, including SLC gaps (U.S. Geological Survey, 2016), snow, cloud/cloud shadow, e.g., a high value for cloud or cirrus confidence, or with a surface temperature uncertainty greater or equal to 6 Kelvin. For model calibration and validation purposes, a single pixel near each measurement site was extracted.

2.3 Estimating daily net radiation and ground heat flux from Landsat observations

We estimated surface net radiation ($R_n$) at a daily time scale as shown in Equation 2,

$$R_n = (1 - \text{Albedo}) \times S_1 + [\varepsilon_s \sigma T_a^4 - (m \times (\varepsilon_s \sigma T_s^4)) + b]$$

where $S_1$ and $T_a$ are daily mean all-sky incoming solar radiation and daily mean air temperature ($T_a$), from CIMIS-spatial (Hart et al., 2009), $\varepsilon$ represents all-sky emissivity for the atmosphere, $\varepsilon_s$ surface emissivity, $T_s$ for instantaneous land surface temperature in Kelvin from CIMIS surface temperature product (Figure 2). Two parameters ($m$ and $b$) to be optimized were included here to convert the estimated instantaneous outgoing longwave radiation to daily mean values.

Shortwave surface albedo was derived at 30m from the Landsat surface reflectance using the narrowband to broadband conversion coefficients (Liang, 2001).

To calculate the incoming longwave radiation, the approach of Jin et al., (2011) was followed to estimate all-sky emissivity ($\varepsilon$) as a function of cloud fraction ($F$) and clear sky emissivity ($\varepsilon_{clr}$), $\varepsilon = (1 + \mu \times F^v \times \varepsilon_{clr})$. Cloud fraction is computed by subtracting 1 by the ratio of Spatial-CIMIS’s all-sky and clear-sky incoming solar radiation. In this study, the parameters $\mu = 0.242$ and $V = 0.583$ (Duarte et al., 2006) were used. $\varepsilon_{clr}$ was estimated with CIMIS-spatial’s dew point temperature.

We developed a different approach from Jin et al. (2011) to estimate the daily mean outgoing longwave radiation from the land surface, considering there is only one Landsat overpassing time around 10:30 am as compared to four from MODIS Terra and Aqua. The instantaneous outgoing longwave radiation was first estimated with Landsat land surface temperature and surface emissivity product (Cook, 2014), and then transformed to daily mean values through a linear relationship, as shown by two parameters ($m$ and $b$) in Equation 2. We
optimized this linear function, with the field measurements of half-hourly longwave radiation, only available from sites 1 and 13 from the training dataset during clear-sky overpassing days. A piecewise linear regression was used from the Shape Language Modeling toolbox (D’Errico, 2020), based on the site measurements of half-hourly longwave radiation, available from sites 1 and 13 from the training dataset during clear-sky overpassing days.

Ground heat flux (G) was estimated from an empirical relationship of $R_n$ and fractions of vegetation and soil ($f_{veg}$, $f_{soil}$) as described in Jin et al., (2011), where $f_{soil} = e^{-0.4 LAI}$, $f_{veg} = 1 - f_{soil}$, and $G = -4.6144 + (0.0496 f_{veg} + 0.1048 f_{soil}) R_n$. We adopted PT-JPL’s approach in estimating LAI using normalized difference vegetation index (NDVI), where $LAI = -\log(1 - (NDVI - 0.05)) / 0.3$ (Fisher et al., 2008). NDVI, calculated based on the Landsat reflectance at red and near-infrared, has been widely used as an indicator of green vegetation cover and biomass (Tucker, 1979).

### 2.4 Actual Priestley Taylor coefficient optimization

The actual Priestley Taylor coefficient $PT_a$ in Equation (1) was parameterized as an empirical function of LAI, moisture condition, and air temperature, as shown in Eq. 3,

$$PT_a = (A - C * e^{-B * LAI}) * (D * (NDMI + 1) + E) * f(T_a)$$

where A, B, C, D, E are parameters to be optimized, and $f(T_a)$ was a relatively weak temperature control, due to the warm weather in the valley, i.e., lowered to 0.05 only if $T_a$ is lower than -5 C, otherwise having the value of 1 (Jin et al., 2011). The soil moisture constraint was replaced with a remote sensing-based normalized difference moisture index (NDMI) (Gao, 1996; Ji et al., 2011), due to the lack of volumetric soil moisture data and the practical consideration in implementing a water balance model as used in Jin et al. (2011) study. NDMI is calculated from Landsat reflectance at two bands, the near-infrared centered at 0.86μm and the shortwave infrared band centered at 1.61μm. It has been used as an indicator of plant water conditions due to strong absorption by liquid water in the shortwave infrared region (Gao, 1996; Ji et al., 2011).

The five parameters (A, B, C, D, E) in Eq. 3 were optimized with daily data from field observations. For each crop type that had field measurements, we used the non-linear least-squares optimization procedure in MatLab Optimization Toolbox (MATLAB, 2018), to minimize the difference between the predicted actual Priestley Taylor coefficients and those derived from the daily means of field measurements. We randomly split the measurement data during all days when Landsat overpassing days were not obscured by cloud, into training (70%) and testing subsets (30%) for each site. To test the robustness of the algorithm, we repeated this random selection procedure 1000 times. For crop types that had seven stations, i.e., alfalfa and corn, modified k-fold cross-validation was further performed, with a “leave-two-out” approach, where all data from two stations was reserved for independent testing at a time, while the rest used for calibration.

To estimate evapotranspiration for crop types where field measurements were not available, a generalized optimization was also developed, where the data were pooled from all crops together for one universal set of optimization. This was done by randomly selecting a maximum of 50 points for each crop from the training dataset mentioned above, due to the unevenly available sample size among crop types. Similar to the crop-specific optimization, we also conducted two types of cross-validations to test the robustness of the generalized method.
2.5 Daily evapotranspiration estimation

During cloud-free days with Landsat overpasses, Landsat-derived LAI and NDMI were fed into Eq. 3 to estimate the actual Priestley Taylor coefficient for each pixel, which was then combined with available energy (Rn - G) to estimate daily evapotranspiration (Figure 2). For days between Landsat overpasses (or pixels) without valid or high-quality values such as cloudy days or over scan-line corrector data gaps (Storey et al., 2005), a temporal interpolation approach was adopted (Allen et al., 2007; He et al., 2017). First, daily evapotranspiration estimates, during the adjacent clear sky Landsat days and within ± 2 months search window, were divided by the concurrent Spatial-CIMIS daily reference evapotranspiration to derive the fraction of reference evapotranspiration (EToF). A shape-preserving piecewise cubic interpolation was applied to this discrete time series of EToF to obtain a continuous time series of daily EToF. We set a requirement of a minimum of 2 valid observations within the search window for a robust interpolation. This temporal interpolation was needed mostly during rainy season in winter and early spring in California, an off-season for the majority of the crops. Finally, daily evapotranspiration for missing days was estimated as a product of the interpolated EToF and Spatial-CIMIS reference evapotranspiration.

2.6 Validation and uncertainty analysis

The uncertainty of Rn, actual Priestley Taylor coefficient, and daily evapotranspiration, directly estimated by our approach during clear-sky Landsat overpass days, was evaluated by comparison with the field measurements in the training, testing, and the full datasets, respectively. Four metrics were used to evaluate the method performance, including $R^2$, $1 - \frac{\sum_{i=1}^{n}(P_i-M_i)^2}{\sum_{i=1}^{n}(M_i-M)^2}$, also known as Nash-Sutcliffe model efficiency coefficient; root mean squared error (RMSE), bias, where $Bias = \frac{\sum_{i=1}^{n}(P_i-M_i)}{n}$, and relative mean absolute difference (RMAD), $RMAD(\%) = \frac{\sum_{i=1}^{n}|P_i-M_i|}{\sum_{i=1}^{n}|M_i|} \times 100\%$, where $P$ and $M$ are remote sensing estimates and field measurements respectively. We also summarized the statistics of the results from the cross-validations, including the variance of the optimized parameters in Eq. 3, and the corresponding estimates of Priestley Taylor coefficient and evapotranspiration values.

Values of daily evapotranspiration that involved temporal interpolation were further evaluated during non-Landsat overpassing or cloudy days. An evaluation was also done for continuous daily, weekly, and monthly evapotranspiration estimates. The uncertainty introduced by the generalized optimization was also analyzed.

To compare our refined approach, hereafter referred to as PT-UCD, with previous Priestley Taylor approaches, we calculated the actual Priestley Taylor coefficients and evapotranspiration for all sites during clear-sky Landsat overpass days, following Jin et al. (2011)’s and Fisher et al. (2008)’s methods, hereafter referred as PT-0 and PT-JPL, respectively. All methods were driven by the same forcing data here, e.g., Rn, CIMIS-Spatial meteorological data, to evaluate the difference caused by the methods themselves. PT-0 did not apply the soil-moisture constraint for irrigated crops, and therefore its actual Priestley Taylor coefficients equation can be simplified as $\alpha = 1.22(1 - e^{-3.48LAI}) \times f(T_0)$. To estimate PT-JPL evapotranspiration using Landsat time series data, we modified PT-JPL’s publicly accessible code (Fisher, 2008), which was written to estimates spatial evapotranspiration using monthly Advanced Very High Resolution Radiometer and The International Satellite Land Surface Climatology Project, Initiative II data (See Text S1 for details).
2.7 Regional evapotranspiration patterns

The Priestley Taylor method optimized here was applied over the whole California Central Valley to estimate crop evapotranspiration during the 2014 and 2016 water years. The crop-specific actual Priestley Taylor coefficient parameterization results were used for daily averaged evapotranspiration estimation over alfalfa, almond, corn, citrus, pasture, and rice areas during Landsat overpassing days. For remaining crop types, including but not limited to grapes, walnut, pistachio, tomatoes, wheat, and cotton, where no field evapotranspiration data were available for crop-specific optimization, the generalized actual Priestley Taylor coefficient parameterizations was applied. Temporal interpolation was applied to derive a complete time series of daily evapotranspiration for each Landsat pixel. For each month, an EToF pixel is interpolated only if there are at least two estimates on clear-sky overpassing days within ±2-month moving time window; the un-interpolated pixels were gap-filled by multiplying daily reference evapotranspiration by EToF averaged by corresponding month and crop within each Landsat Analysis Ready Data tile. Daily evapotranspiration estimates were further averaged to annual time scales to analyze the regional patterns. Evapotranspiration was summarized for each crop type and compared the differences among crops by evaluating the annual evapotranspiration, reference evapotranspiration, and EToF. Specifically, the per-area water consumptive use average was computed by dividing the sum of annual evapotranspiration by crop area over non-gap-filled pixels, while total consumptive use was computed over all crop area pixels.

We further summarized annual evapotranspiration by GSA boundaries to provide agricultural water use information for water planning. This was achieved by quantifying annual water use and variability for each planning area and compared across areas. We also analyzed the association of water use with corresponding land use, Rn, actual Priestley Taylor coefficient, EToF, and reference evapotranspiration, to understand what contributed to water use differences among GSAs. While GSAs manage local groundwater resources, DWR oversees water resources regionally by water planning area. We summarized our annual crop evapotranspiration estimates by water planning areas in the Central Valley and compared them with DWR’s estimates for the water year 2014.

3 Results

3.1 Net Radiation and Available Energy

The $R_n$ estimated from our approach captured similar seasonal dynamics as observed by field measurements (Figure 3). For example, estimated $R_n$ was close to 20 Wm$^{-2}$ in January, increased continuously to 200 Wm$^{-2}$ in June over an example alfalfa site (#2) and corn site (#14), and then started to decrease in July, reaching around 75 Wm$^{-2}$ in October. An overestimation of $R_n$ was found over Rice, mostly when the fields were flooded, e.g., during the growing season from April to October (Knox et al., 2016) and in winter. There is also a recurring underestimation of corn’s $R_n$ in between April and July for both sites 13 and 14, but not site 15 and 19.

Overall the remote sensing-derived $R_n$ agreed well with $R_n$ measurements across all sites and overpassing days, with an overall $R^2$ of 0.74, an RMSE of 25.8 Wm$^{-2}$, and an RMAD of 14.3% for the testing subset (Figure 4b). Only a small bias of 1.70 Wm$^{-2}$ was found. The agreement was similar to the training subset, with an $R^2$ of 0.70 and an RMSE of 27.4 Wm$^{-2}$ (Figure 4a). The uncertainties of $R_n$ estimate varied by crop type, with the highest accuracy

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found for alfalfa, citrus, and pasture while relatively larger uncertainty for rice. Note that both
evaluations using training and testing datasets were informative about the model’s uncertainty
because there was only one calibration for our Rn estimate, converting the instantaneous to daily
outgoing longwave radiation on clear-sky Landsat overpassing days, which used only two sites
(n = 94) in the training dataset.

3.2 Actual Priestley Taylor coefficients

The seasonal dynamics of the actual Priestley Taylor coefficient typically followed the
plant growth curve, as shown by the values derived from both the field measurements and
satellite observations (Figure 5). For example, the actual Priestley Taylor coefficients of alfalfa
frequently fluctuated from 0.5 to 1.5, likely due to the multiple cuttings throughout the growing
season, as shown by the similar variations in LAI (Figure 5a). Field measurements showed a
substantial seasonal variation in the actual Priestley Taylor coefficient for the corn and rice sites,
e.g., with towering peaks in summer growing season, a relatively small peak in spring, and much
lower values in between fall and winter (Figures 5b, c).

In general, the remote sensing-derived actual Priestley Taylor coefficients, from the crop-
specific optimization, could explain 56% of the variance observed across sites and time periods,
with an RMSE and RMAD of 0.23 and 17.7%, when compared with the field-based estimates
over the testing dataset (Figure S1). For the generalized optimization, the uncertainties of actual
Priestley Taylor coefficient estimates increased slightly (e.g., R² = 0.50, RMSE = 0.26, and
RMAD = 21.64%). Among crop types, both crop-specific and generalized actual Priestley Taylor
coefficient estimation performs best for almond (RMAD = 8.2% and 9.4%). The performance of
the crop-specific actual Priestley Taylor coefficient is significantly better than the generalized
actual Priestley Taylor coefficients for corn and citrus.

The actual Priestley Taylor coefficient estimates showed significant improvement when
compared to those derived from PT-0, which only captured small seasonal variation (25%) and
had a higher bias of 0.24 and a larger RMAD of 34.7% over the irrigated cropland in the valley
(Figure 5). In contrast, PT-JPL estimates showed a reasonable seasonal pattern for alfalfa and
corn (Figures 5a, b), although it was not calibrated for any land cover type (RMSE = 0.32, and
RMAD = 27.3%). Across all sites, the crop-specific PT-UCD showed an overall improvement
over PT-JPL, as shown by the empirical cumulative distribution function of the absolute errors
when compared to both testing and independent testing data (Figure 6a). For example, 88% of
testing samples had an absolute error were below 0.30 from crop-specific PTUCD estimates,
compared to 62% and 59% from PT-JPL and PT-0 estimates, respectively. The generalized PT-
UCD performed only slightly better than PT-JPL (Figure 6a).

Two types of cross-validation testing further showed the optimization of the parameters
in Eq. 3 for estimating the actual Priestley Taylor coefficient was reasonably robust. The
distribution of the estimated parameters showed a very small variance, for the majority of the
crops and the generalized optimization (Table S2). One exception was Parameter D, which
represented the moisture regulation over the coefficient, for citrus and pasture (typically well-
watered). The estimated actual Priestley Taylor coefficients were shown to be stable among the
repeat and leave-two-out cross-validations (Figure S2-5), with an Inter Quantile Range (IQR) of
RMAD of less than 5% (Figure S2).
3.3 Evapotranspiration estimates

3.3.1 Daily mean evapotranspiration on clear-sky Landsat overpassing days

We found a good agreement between field measurements of evapotranspiration and satellite-based estimates during the clear-sky days with Landsat acquisitions. When evaluated with the testing dataset, both the crop-specific and generalized evapotranspiration models captured the seasonal variability well (Figure 7). Across all sites, the crop-specific evapotranspiration had an $R^2$ of 0.79, RMSE of 0.90 mm day$^{-1}$, and RMAD of 20.5% (Figure 8b). Only a small bias of 0.14 mm day$^{-1}$ was found. When using the generalized actual Priestley Taylor coefficients, slightly higher uncertainties were found, with an $R^2$ of 0.76, RMSE of 0.98 mm day$^{-1}$, and RMAD of 23.1% (Figure 8d).

The performance of evapotranspiration estimates varied by crop types. When using the crop-specific Priestley Taylor optimization, the RMSE and RMAD ranged from 0.68 to 1.34 mm day$^{-1}$ and 13.3% to 28.4%, based on the comparison with the testing dataset (Table 2). The best performance was found for alfalfa, citrus, and pasture sites, while the weakest performance in rice. The generalized approach also performed the best for alfalfa and citrus and performed the poorest for rice and corn (Figure 8, S3, Table S4). The leave two out cross-validation showed relatively small differences in RMSEs of daily ET estimates from site to site (Figure S4), e.g., 0.7 mm day$^{-1}$ in alfalfa site #6 vs. 0.9 mm day$^{-1}$ in site 5 based on the results from alfalfa-specific optimization, and 0.7 to 1.2 mm day$^{-1}$ among the corn sites.

Crop-specific PT-UCD showed an improvement over PT-0, PT-JPL, and generalized PT-UCD. About 80% of crop-specific evapotranspiration estimates in the testing and independent dataset had an error of less than 1 mm day$^{-1}$, as shown by the empirical cumulative distribution functions of the absolute errors between the daily crop-specific evapotranspiration estimates and field measurements (Figure 6b). In contrast, both generalized PT-UCD and PT-JPL appeared to perform similarly, i.e., about 70-76% of samples had an evapotranspiration error less than 1 mm day$^{-1}$, and about 85-90% less than 1.5 mm day$^{-1}$. However, for the PT-0 evapotranspiration estimates, only 55% and 70% of samples had an error less than 1 mm day$^{-1}$ and 1.5 mm day$^{-1}$, respectively.

3.3.2 Continuous evapotranspiration estimates

The interpolation of EToF from adjacent overpassing days introduced a small overall uncertainty in daily evapotranspiration estimates, e.g., RMSE increased by 0.10 – 0.17 mm day$^{-1}$ and decreased $R^2$ by 0 – 0.08 when estimating evapotranspiration for alfalfa, citrus, corn, and pasture (Table 2, S6). When further aggregated to weekly and monthly time scales, the satellite-derived evapotranspiration estimates agreed better with those from the field measurements (Figures S8c, d). For example, across all sites, $R^2$ was increased to 0.83 and 0.88, and RMSE reduced to 0.79 and 0.65 mm day$^{-1}$, respectively, for weekly and monthly evapotranspiration values based on the crop-specific Priestley Taylor optimization.

3.4 Evapotranspiration patterns and dynamics over the Central Valley

We estimated a total of 19.9 and 21.8 tera-liter of water consumption via evapotranspiration over the agricultural land in California’s Central Valley in 2014 (24.1 thousand km$^2$) and 2016 water years (25.8 thousand km$^2$), respectively (Figures 9c, d). In the water year 2014, the top eight crop types accounted for 75% of total crop consumptive water use in the valley, including almond (22.2%), rice (10.2%), grapes (9.4%), alfalfa (9.4%), corn...
(8.9%), walnuts (7.3%), pistachios (4.1%), and tomatoes (3.6%) (Figures 10a, b). Overall, rice was the second largest water consumer after almonds, although it only used 7.6% of cropland, due to its highest annual evapotranspiration rate of 1109 (± 85) mm yr⁻¹, on a per unit area basis (Figure 10c). Pasture, walnut, almonds, citrus, and alfalfa also had relatively high evapotranspiration rate (greater than 800 mm yr⁻¹) (Figure 10c) (Table S7). In contrast, wheat consumed the least amount of water per area (561 ± 228 mm yr⁻¹) among major crops, whereas pistachio, tomato, corn, cotton, and grapes had a moderate evapotranspiration rate between 600 and 800 mm yr⁻¹. As the second and third largest cropland use in the valley, grapes, and corn used less total water than rice but similar water with alfalfa, walnuts, and fruits.

Across the valley, the mean annual evapotranspiration rate varied by 35% (820 ± 290 mm yr⁻¹) over all agricultural pixels in 2014, mostly due to the diversity of the crop types. We also found high variability of evapotranspiration rate within each crop type, especially over orchards such as almond, pistachios, and walnut, with a CV higher than 20% (Table S7), most likely due to differences in planting density, age, canopy structures and stressors among orchards (Ferguson et al., 2005). For example, the almond evapotranspiration rate varied by 34% (964 ± 328 mm yr⁻¹), and the rate for pistachio varied by 59% (592 ± 352 mm yr⁻¹) in 2014. Wheat also had a very high variability (CV: 41%), different from other annual crops, which typically had a much lower variation of evapotranspiration rate than perennial crops. For all major crop categories, the difference in CV between 2014 and 2016 was less than 7.2%.

Compared to 2014, total crop consumptive water use increased by 9.6% in 2016 (Figure 10b), with an evapotranspiration rate of 856 (± 306) mm yr⁻¹, although the reference evapotranspiration from Spatial-CIMIS decreased by 4%. This increase in evapotranspiration was mostly caused by land-use changes with higher irrigated areas and crops with higher averaged water consumptive use (Figure 10a, c). Total irrigated agriculture land use increased by 7.0% in 2016, partly due to a 2370 km² land-use conversion from fallow/idle lands in 2014 to cropland in 2016. A large portion of fallow land conversion grew rice (18%), wheat (16%), and perennial (9.7%) crops in 2016, leading to an increase of total water use by 1.3 tera-liters.

Another major land-use change was the conversion from annual crops (386 km²) to high water demand orchards, including almonds, walnuts, citrus, or grape in 2016, accounting for 1.5% of 2016 cropland and decreasing water use by 0.07 tera-liters due to the low evapotranspiration rate of young orchards.

3.5 Evapotranspiration difference among GSAs

Across a total of 134 GSAs in the valley, the annual mean per-crop-area evapotranspiration rate was averaged at 803 (±154) mm yr⁻¹ in the 2014 water year, and it varied significantly among GSAs, with a CV of 19% and IQR of 173 mm yr⁻¹ (Figure 11a). Most GSAs with a high ET rate (>1000 mm yr⁻¹) were located in the north-eastern Central Valley, such as the Vina, Yuba Water Agency, and Southern Sutter Water District GSAs in Butte, Sutter, and Yuba counties. On the other hand, most GSAs with ET rates lower than 700 mm yr⁻¹ were scattered across the south-western Central Valley, e.g. East Contra Costa Irrigation District GSA in the Contra Costa County, City of Tracy GSA, and the City of Stockton GSA in the San Joaquin County, El Rico GSA in the Kings county, and Pleasant Valley GSA and Westlands Water District GSA in the western part of Fresno county.

Variability of evapotranspiration rate among GSAs was primarily driven by non-meteorological drivers. Across GSAs, we found that the evapotranspiration rate highly correlated with net radiation (r=0.92) and actual Priestley Taylor coefficient (r=0.88) (Figures 11b, c).
Many of these factors were regulated by land-use types, vegetation cover, and plant water stress status. The annual ETo (1498 ± 74 mm yr⁻¹) could only explain 4.5% of the variation in annual evapotranspiration rate among GSAs (Figure S9a). In contrast, EToF was the dominant driver of evapotranspiration rate variability among GSAs (r=0.96) (Figure 11d), mostly driven by crop types, e.g. rice with EToF of 0.61 (± 0.06), tomato 0.33 (± 0.07), almond 0.52 (± 0.11), and pistachio 0.37 (± 0.16) summarized at the GSA scale.

Even for the same crop type, EToF varied significantly among GSAs for some tree crops and wheat 0.36 (± 0.09). The average almond EToF (Figure 12c), for example, ranged from 0.25 in the City of Tracy GSA in Tracy county to 0.75 in Rock Creek Reclamation District GSA in Chico county. Pistachio’s EToF was much lower in the majority of the western San Joaquin Valley areas (Figure 12d), probably due to the plant stress caused by salinity (Hanak et al., 2017; Jin et al., 2018; Letey, 2000). Citrus EToF (0.57 ± 0.11) had an IQR of 0.16 at the GSA scale. In contrast, the mean EToF showed much smaller variation among GSAs for the majority of annual crops such as alfalfa 0.54 (± 0.08) (Figure 12a), rice 0.61 (± 0.06) (Figure 12b), pasture 0.57 (± 0.09), and cotton 0.35 (± 0.05).

Within each GSA, the annual EToF also showed large spatial variation, with a mean CV of 31% (± 12%) across all agricultural fields; Some GSAs with a lower evapotranspiration rate had the highest variability (Figures S9c, 11a), mostly located at and around the Westlands Water District region in the western-Fresno and Kings county. In addition to crop diversity within each GSA, significant variation of EToF was also found for each tree crop type, such as almonds and pistachios (Figures 12g, h), with IQRs of CVs among GSAs greater than 15%. For examples, the CVs of pistachio EToF within each GSA had a mean of 34% (± 20%) and an IQR of 33% across GSAs, with the largest within-GSA variation found in the Central Delta-Mendota GSA; In the Southeast Kings GSA, CV of Pistachio EToF is 29%, much lower than its neighbor, Tri-County Water Authority GSA (64%). Other types with highly variable EToF (CV of 18%± 10%) included almond, citrus, walnut, and wheat. In contrast, EToF was more homogeneous within GSA for alfalfa with a mean CV of 17% (± 7%) and IQR of 6% (Figure 12e), and rice (Figure 12f).

About 39 GSAs had more than 60% of agricultural land areas planted with perennial crops including almond, pistachio, citrus, walnut, and grape in 2014, which accounted for 76% of total agricultural water use by these GSAs and 27% of Central Valley’s total agricultural water use in 2014 (Figure 13b). These GSAs will likely face greater vulnerability to prolonged drought due to the high cost of fallowing productive orchards. When dividing the total consumptive use of perennial crops by the GSA area (Figure 13c), we found that some small and medium-size GSAs, such as Delano-Earlimart Irrigation District GSA, Madera Water District GSA, and New Stone Water District GSA, will need to reserve a much greater depth (400 mm yr⁻¹) of groundwater storage to maintain the orchards during drought.

4 Discussion

4.1 Site-level performance

Our study showed that the semi-empirical Priestley Taylor algorithm, when calibrated with ground measurement data over diverse crop types and driven by Landsat Analysis Ready data, improved the accuracy of the older 1 km MODIS-driven PT-0 model (Jin et al. 2011). The crop-specific Priestley Taylor optimization performed consistently between the testing and independent datasets, and slightly better than the PT-JPL method (Fisher et al. 2008) (Figure 6,
The generalized Priestley Taylor optimization had a similar overall performance with PT-JPL when driven by the same input data. However, relatively larger uncertainties were found during non-growing seasons, from November to March, when the evapotranspiration rate was relatively low. This was partly due to the limited field measurements data during winter and early spring for optimizing the sensitivity of actual Priestley Taylor coefficients to the moisture content. Moreover, our Priestley Taylor approach does not separate soil evaporation and plant transpiration. This introduces uncertainty in evapotranspiration estimates during non-growing seasons when evapotranspiration is mainly driven by evaporation from the soil due to minimal canopy coverage or leaf area. For example, we did find that PT-JPL better captured the peak of the actual Priestley Taylor coefficient for the corn site during the dormant season (Figure 5b), when PT-JPL’s estimates showed that soil evaporation was the most significant component.

The uncertainty of our refined Priestley Taylor approach here is similar to the DisALEXI model, as shown by the report from the Sacramento San Joaquin Delta inter-comparison project (Medellin-Azuara, J. et al., 2018). For additional reference, Anderson et al., (2018) reported that DisALEXI had an RMSE of 1.09 mm day⁻¹ at site number 1 (alfalfa) and 1.24 mm day⁻¹ at site number 24 (rice) when compared to daily measurements. Being a process-based model, DisALEXI does not depend on land use maps and field measurements for calibration once validated. The semi-empirical Priestley Taylor approach, however, has the advantage of easy implementation, compared to other more sophisticated and computationally more expensive approaches.

4.2 Regional evapotranspiration estimates

At a regional scale, the annual mean values of per-area water use of major crop types in the Central Valley (Table S7) estimated here are generally within the ranges reported in the literature (Burt et al., 2002; California Department of Water Resources, 2019a; Schauer & Senay, 2019). For example, DWR’s water portfolio and balances dataset, as part of DWR’s 2018 Water Plan, reports that water requirement by corn ranges from 390-835 mm yr⁻¹ in 2014 across sub-regions (Detailed Analysis Units by County) of all planning areas in the Central Valley (California Department of Water Resources, 2019a,d). Burt et al., (2002) estimated that corn in the Central Valley conventionally used 813 mm yr⁻¹ in a typical precipitation year. Our regional average of corn evapotranspiration (705 mm yr⁻¹) was 16% more than DWR’s average corn water requirement over planning areas (604 mm yr⁻¹). Larger differences were found for alfalfa, pasture, wheat, almonds, pistachio, and vineyard, for which our regional averages were 30-65% lower or higher than DWR’s values. Over all 30-m pixels of agricultural lands in Central Valley (~25 thousand km²), the average annual evapotranspiration rate, estimated here, is higher than the estimates over the whole Central Valley (~47 thousand km²) by the BESS biophysical process-based model forced with 1km satellite observations (Baldocchi et al., 2019). The discrepancy is likely due to the scale effect (Wu & Li, 2009) and differences in land cover maps. Larger pixels likely contain other land use areas such as fallow, urban, water, and natural vegetation. Nonetheless, our estimates in 2014 align with the values reported in Schauer & Senay (2019) based on the SSEBop remote sensing evapotranspiration model driven by Landsat thermal data.

Our estimation of 19.6 tera-liter water consumption in 2014 was equivalent to 74% of DWR’s estimate (26.2 tera-liter) over all planning areas within the Central Valley, which was derived from CalSIMETAW(Orang et al., 2013). Among the planning areas, the discrepancies ranged from – 53.4 to -18.5%, with the most significant disagreement (<-50%) occurring in the

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Similarly, previous studies in the Sacramento San Joaquin Delta, a subset of the Central Valley, showed that remote sensing estimates are lower than CALSIMETAW’s estimates by 6 to 24% (Medellín-Azuara, J. et al., 2018; Paul et al., 2017). Over this Delta area, our crop-specific Priestley Taylor method in this study estimated 1.20 tera-liter in 2016, very similar to the DISALEXI’s estimate of 1.16 tera-liter in the water year 2016; both were about 80% of CalSIMETAW’s estimates of 1.49 tera-liters, based on the published data summary table in Medellín-Azuara et al., 2018.

Two factors may have caused the discrepancy in regional estimates between PT-UCD and CalSIMETAW. First, CalSIMETAW’s crop coefficient approach implemented at a regional scale may overestimate actual evapotranspiration, because it didn’t account for the impacts of planting variabilities such as orchard age distribution and planting density, field conditions such as salinity and disease, and crop management like deficit irrigation. Secondly, the land-use map used by CALSIMETAW was different from the DWR’s land use map that we used here. For example, CALSIMETAW estimated 13.7, 23.9, and 23.9 km² of corn, alfalfa, and pasture in PA 704 in 2014, in contrast to our DWR’s map-based estimates of 18.2, 21.4, and 9.3 km² of corn, alfalfa, and pasture.

4.3 Implications for water resource management

Currently, California’s Groundwater Sustainability Agencies (GSAs) employ various approaches to estimate evapotranspiration in their water budget accounting and management plan development, causing systematic inconsistencies among GSAs. For example, the Olcese GSA near Bakersfield estimates monthly evapotranspiration from 1993 to 2015 using the METRIC method version by the Irrigation Training & Research Center (ITRC) at the California Polytechnic State University; North Kings GSA uses CA DWR’s crop coefficients to estimate annual evapotranspiration rate over detailed analysis units from 1998 to 2010, while the Delano and Yuba GSAs use crop coefficients published by ITRC in 2003 and derived from a SEBAL-based evapotranspiration map in 2009, respectively. Our study shows that the fractional of reference ET (EToF), or similarly crop coefficients, for most crops, varies spatially across and even within GSAs, and for some crops, EToF changes considerably between years. More consistent estimates with known uncertainty from a calibrated or thoroughly evaluated approach are needed to ensure consistent quantitative information for data-driven decisions for water planning. Our optimized Priestley Taylor approach driven by remote sensing observations provides an efficient way to capture both spatial heterogeneity and temporal dynamics of water balance.

In particular, we found that orchards and wheat generally had a greater spatial variability of evapotranspiration and crop coefficients than other major crop types, across the Central Valley, within, and among GSAs. Age distribution and other stressors such as salinity likely contributed to such evapotranspiration variability for tree crops (Jin et al., 2018; UC-ANR, 2008, 2018). Among three major nut tree crops, pistachio had the lowest mean annual evapotranspiration rate (592 ± 352 mm yr⁻¹ in 2014), followed by walnut (981 ± 291 mm yr⁻¹ in 2014) and almond (964 ± 328 mm yr⁻¹ in 2014). Coincidentally, 26% of pistachio acres in 2014, 18% of walnut in 2015, and 15% of almond in 2014 across California were non-bearing orchards (Administrative Committee for Pistachios, 2020; California Department of Food and Agriculture, 2020; US Department of Agriculture, 2020).
The high variability of wheat water use is likely due to cultivar and end-use for the crop (Fulton et al., 2006).

The rapid expansion of perennial crop acreage in the past two decades raises concerns about increasing and hardening water demand (Hanak et al., 2017; Johnson & Cody, 2015; Mall & Herman, 2019). As perennial crops have a sizeable initial investment cost, following perennial crops during drought results in greater economic losses than fallowing annual crops. Therefore, GSAs with a high percentage of water use by perennials (Figure 13b) will likely experience challenges in implementing sustainability management during drought. To minimize economic loss during a severe multi-year drought, a large buffer between the sustainability threshold and actual water level should be maintained for those GSAs, i.e., to prepare for a five-year drought with conditions like 2014’s, a groundwater buffer should be aimed with an approximate depth at least five times the values shown in Figure 13c, after accounting for surface water availability and total porosity of the aquifer.

The NASA ECOSTRESS mission (Fisher et al., 2020) has an ongoing partnership with USDA, states of California, Florida, and Iowa, and many water districts on using remote sensing evapotranspiration to better inform water resources management (Anderson et al., 2021). Anderson et al., 2021 also find that the improved temporal frequency of ECOSTRESS resulted in improved evapotranspiration estimates (RMSE improved by 65%) and captured peak growing season during which there were no Landsat acquisitions. The mission adopted both PT-JPL and DisALEXI to map evapotranspiration (L3 products), which serves as the basis for the derive L4 products, such as Water Use Efficiency (the ratio of gross primary productivity to actual evapotranspiration) and Evaporative Stress Index (the ratio of actual to reference evapotranspiration). ECOSTRESS also operationally provides the Priestley Taylor potential evapotranspiration product, which has been demonstrated to be useful for water management agencies for spatial estimates of reference evapotranspiration (Kohli et al., 2020). Our model evaluation work suggests that PT-JPL’s evapotranspiration estimates could potentially be further improved over irrigated croplands in agricultural regions with ample evapotranspiration measures over diverse crop types. On the other hand, our spatial estimates show that EToF, which is analogous to ECOSTRESS’S Evaporative Stress Index, varies by crop types and within the Central Valley for the same crop types. Users of the Evaporative Stress Index product over Central Valley should also account for the threshold of water stress dependencies and variability by crop types and other factors such as orchard age and salinity.

4.4 Model limitations and potential improvement

Although overall our refined evapotranspiration estimation approach here has similar performance to that in more complex models such as DisALEXI (Anderson et al., 2018; Medellin-Azuara, J. et al., 2018), there are still a few factors that can cause errors in our estimates. PT-UCD is a single-source approach. We noticed an overestimation of rice net radiation when the field was flooded, probably due to the challenges posed by heat storage in the water column and the effect of a wet surface, and an underestimation of net radiation over the two Ameriflux corn sites between every April and July. We also found a larger uncertainty in estimating actual Priestley Taylor coefficients for corn during the dormant season between January and April. Although evapotranspiration during the non-growing season accounts for a small fraction of total annual water use, explicit consideration of soil and plant components of energy balance is expected to improve the accuracy of evapotranspiration estimates.
The errors of field measurements of evapotranspiration can propagate to the optimization. The energy balance closure issue, for example, has been well recognized for the eddy covariance measurements (Twine et al., 2000; Wilson et al., 2002). Baldocchi et al., 2016 and Eichelmann et al., 2018, for example, conducted several analyses at AmeriFlux sites in the Sacramento San Joaquin Delta and found a closure of 79.3% in an alfalfa site, i.e., the ratio of the sum of sensible and latent heat flux over the residual between net radiation and ground heat flux and storage, and 71% in a rice site. Their study suggests that incomplete storage calculation, rather than underestimation of eddy covariance measurements of fluxes, plays a major role in the lack of energy balance closure for their sites in the Sacramento San Joaquin Delta. Therefore, we did not perform the correction for eddy covariance measurement sites from our dataset.

Additional uncertainties can be introduced due to the varying footprint size of the flux towers and the scale mismatch between 30m ET and tower measurements. We compared the annual evapotranspiration value extracted from a single Landsat pixel collocated with measurement sites and the mean (and standard deviation) of the values within the surrounding larger areas (Table S6). The differences in estimates between a 90m by 90m window and the single-pixel were small, with the largest difference of -3.2% occurred at site 18 in 2016. The differences generally increased with a larger footprint, depending on the heterogeneity of the areas, e.g., over a 510 m by 510 m window at site 16 (corn) in 2016 had the largest mean relative difference of -10.3% from the corresponding center pixel value. Other studies also suggested that a rigorous footprint approximation is needed in future studies to make a fair comparison with field measurements (Anderson et al., 2018). For example, (Kljun et al., 2015)’s flux footprint model could be implemented at flux tower sites to determine the weight and extent of the pixel window.

Compared to those driven by MODIS with daily revisiting (Jin et al., 2011), the evapotranspiration estimates derived from Landsat have the benefit of capturing finer spatial details, which is critical for water use assessment over a heterogeneous landscape. Landsat’s 16-day repeating cycle, however, can potentially lead to uncertainty in water use monitoring, especially during the rainy season or during the rapid plant growth and senescence stages. In this study, the uncertainty due to the temporal interpolation of the missing days was found minimal overall, likely because the rainy season coincides with dormancy or the very beginning or ending of the crop growth for the majority of crops in Central Valley due to its Mediterranean climate (Figure S6, Table 2, S5). There were situations, e.g., right after irrigation or right after harvesting for crops that undergo multiple harvests (alfalfa), when a relatively large error was introduced from the estimates interpolated from observations a few days apart. Future work is also needed to increase the temporal resolution of water use estimate by fusing Sentinel 2 A&B satellite observations every 5 days with Landsat data. A sophisticated data fusion technique will also improve the accuracy of evapotranspiration monitoring and assessment, by taking advantage of complementary observations from multiple sensors (Anderson et al., 2018).

The robustness of our optimization approach partly relies on the availability of multiple field measurements for diverse crop types across the Central Valley. The automatic workflow developed here allows for a continuous improvement of the optimization accuracy, by taking advantages of the increasingly available crop evapotranspiration measurements with the increased deployment of both research-grade and commercialized surface renewal stations in the state (Clay et al., 2019; Jin et al., 2018; Souto et al., 2019; Xue et al., 2020). Although the Priestley Taylor parameters in this study were tailored for California’s cropland, our data-model integration framework is generalizable to other regions. Once recalibrated and tested with local
field data, the PT-UCD approach can be applied to monitor daily evapotranspiration and assess water use at various scales over regions besides the Central Valley.

5 Conclusion

To estimate agriculture consumptive water use at the field scale in Central Valley, we further refined the semi-empirical Priestley-Taylor approach (Jin et al., 2011), by optimizing the actual Priestley Taylor coefficient as a function of the leaf area index and moisture index with recently compiled ground measurements of evapotranspiration for California’s major crops. The net radiation estimates with Landsat Analysis Ready Data had an RMSE of 25.8 Wm$^{-2}$ and R$^2$ of 0.74. Our study shows that the crop-specific optimization with field data is relatively robust, and the locally optimized approach captured well the regulation of plant growth and moisture stress on latent heat exchange and thus improved the accuracy of evapotranspiration estimation. The daily evapotranspiration, directly estimated during Landsat overpassing days under clear-sky and interpolated for other days, largely agreed well with the field measurements. The validation showed an R$^2$ of 0.78 and an RMSE of 0.95 mm day$^{-1}$. To facilitate the implementation of sustainable groundwater management plan in California, we applied this calibrated approach to the entire Central Valley, one of the world’s most productive agricultural regions, to estimate evapotranspiration at 30m resolution for water years 2014 and 2016 when high accuracy local land use maps are available. The evapotranspiration rate averaged at 820 mm yr$^{-1}$ and had a large spatial variability with a standard deviation of 290 mm yr$^{-1}$ over all agricultural pixels in 2014. Total water use in 2016 increased by 9.6%, as compared to 2014, mostly as a result of land-use conversion (2370 km$^2$) from fallow/idle land in 2014 to cropland in 2016. When aggregated by the GSA boundaries, the annual evapotranspiration rate showed large variation among GSAs, i.e., 803 ±154 mm yr$^{-1}$ and a CV of 19% in 2014. This variability was mostly associated with variation in net radiation and the actual Priestley Taylor coefficient or fraction of reference evapotranspiration, driven by crop type diversity and conditions. Perennial crops such as pistachios and almonds typically had much higher variability within and among GSAs than annual crops.

Our study demonstrates that remotely sensed evapotranspiration estimates can be combined with land use maps to inform water resources planning and management. In particular, we provided more realistic estimates of crop coefficient and evapotranspiration for local water budget accounting, evaluation of crop water use variation within and among GSAs, and highlighted the importance of crop planning, orchard age structure, and other physiological stressors such as salinity on the trajectory of local orchard water use and demand.

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Data

All data needed to reproduce the reported finding or conduct additional spatial analysis is made available on the HydroShare repository on April 10th, 2021 (http://www.hydroshare.org/resource/fe0ea8781ae44c97a7c4a72be17dfd89). The repository dataset (3.12GB) includes a single table containing model input data and daily averaged measurements from 26 sites (training, testing, and independent subsets of this table on Landsat overpassing days are also made available), calibrated parameters, additional evaluation (e.g. validation of the four sub-components of net radiation, and time series for all sites), annual evapotranspiration maps (unit: mm day⁻¹, Gap-filled and non-gap-filled), and summarized statistics for each GSA and DWR’s planning area. Additional intermediate and output data and data visualization will be added to the repository upon request. All input data used by our model can be accessed from the websites as follows: Landsat Analysis Ready Data (https://earthexplorer.usgs.gov/); Spatial-CIMIS (http://cimis.casil.ucdavis.edu/cimis/); 2014 & 2016 Statewide Land-use Map (https://data.cnra.ca.gov/dataset/statewide-crop-mapping).

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Figure and Table Captions

Figure 1. Study area and locations of measurement sites. Also shown are 2014 agricultural land use (a) over California (black line), County boundaries (orange lines), DWR water planning areas (colored base map) (b), and Groundwater Sustainability Agencies (colored base map) (c) in the Central Valley (red line).

Figure 2. The flowchart of estimating daily evapotranspiration at 30 m during clear sky Landsat overpassing day.

Figure 3. Time Series of Rn from daily measurements (blue) and estimates on clear-sky overpassing days (red) over (a) an alfalfa site at Bouldin island (#2); (b) a corn site at Bouldin island (#14) and (c) a rice site at Twitchell Island (#24). The overpassing days of Landsat 7 and 8 are indicated by the vertical dash lines in grey.

Figure 4. Comparison of remote sensing derived Rn vs. field measurements over all sites, from (a) training data (N=447) and (b) testing dataset (N=389).

Figure 5. Time series of crop-specific (red) and generalized (yellow) PT coefficients derived from tower observations and crop-specific remote sensing estimates, over the same three sites as shown in Figure 3. The un-filled symbols on overpassing dates indicate that the data point were previously partitioned into the training dataset. The estimates from the other two PT approaches, namely PT-0 and PT-JPL, were also shown for comparison. The lower panel shows the corresponding LAI and NDMI in green and cyan (lower panel), ranging from 0 to 7 and -1 to 1, respectively. All three Priestley Taylor methods shown here were driven with the same forcing data.

Figure 6. Comparison of empirical cumulative distribution errors among estimated (a) PT coefficient and (b) evapotranspiration using four models on clear-sky Landsat overpassing days. All three Priestley Taylor methods shown here were driven with the same forcing data.

Figure 7. Time Series of crop-specific and interpolated evapotranspiration vs. daily evapotranspiration measurement.

Figure 8. Comparison of remote sensing-derived daily evapotranspiration vs. field measurements on Landsat overpassing days over all sites, based on (a) training data (N=345) and (b) testing dataset (N=149), when using crop specific optimization parameters. The corresponding results from the generalized optimization are shown in (c) and (d).

Figure 9. Spatial distributions of (a, c) land use and (b, d) annual evapotranspiration (m yr⁻¹) over the Central Valley in 2014 and 2016 water years, respectively.

Figure 10. Primary agricultural land use (a) and agricultural consumptive water use (b,c) for each crop type in the Central Valley for 2014 and 2016 water years. Standard deviation of per-area evapotranspiration was also shown (error bars) in (c). Crops are sorted by the combined area in 2014 and 2016. Individual crops with an area less than 690 km² in either 2014 or 2016 were grouped into one of the following three categories, vegetables, other fruits, and others. Others
include flowers, nursery, Christmas tree, miscellaneous deciduous, miscellaneous grain and hay, miscellaneous grasses, safflower, sunflowers, and young perennials.

**Figure 11.** Spatial variation across Groundwater Sustainability Agencies in 2014 for (a) Per-area evapotranspiration rate (m yr$^{-1}$), (b) Net radiation (Wm$^{-2}$day$^{-1}$), (c) Actual Priestley Taylor coefficient, (d) EToF over all cropland averaged by GSAs.

**Figure 12.** Mean (Top) and CV of the fraction of reference evapotranspiration by GSAs in 2014 for alfalfa (a, e), rice (b, f), almond (c, g), and pistachio (d, h).

**Figure 13.** Percent of cropland (a) and of total water consumptive use (a) in 2014 by major perennial crops (Almond, Pistachio, Citrus, Walnut, and Grape) in GSAs, and their total water consumptive use normalized by GSA area size (c).
Table 1. List of in situ measurement stations and evapotranspiration data availability.

| Crop type (site days) | ID | Site | Time Period            | Citation                      |
|----------------------|----|------|------------------------|-------------------------------|
| Alfalfa (3870)       | 1* | US-Tw3: Twitchell Island | May 2013 – Dec. 2015         | Eichelmann et al., 2018; Hemes et al., 2018 |
|                      | 2* | US-Bi1: Bouldin Island  | Aug. 2016 – Sep. 2019        |                               |
|                      | 3‡ | D02: Staten Island     | Mar 2016 – Apr. 2017         |                               |
|                      | 4‡ | D07: Bouldin Island    | Jul. – Nov. 2016             | Medellín-Azuara, J. et al., 2018 |
|                      | 5‡ | D10: Bacon Island      | Jul 2016 – Arp. 2017         |                               |
|                      | 6‡ | D13: Roberts Island    | Jul. 2016 – May 2017         |                               |
|                      | 7‡ | D14: Twitchell Island  | Aug 2016 – Dec. 2016         |                               |
| Almond (1405)        | 8* | A1: Lost Hill          | Jan. 2009 – Dec. 2011        | Couvreur et al., 2016         |
| Citrus (486)         | 9† | E1: 2 years old        | Jul. – Aug. 2004             | Consoli et al., 2006          |
|                      | 10†| E2: 4 years old        |                               |                               |
|                      | 11†| E3: 15 years old       |                               |                               |
|                      | 12†| E4: 34+ years old      | Feb. 2003 – Aug. 2004        |                               |
| Corn (1657)          | 13*| US-Tw2: Twitchell Island| May 2012 – May 2013          | Hemes et al., 2018            |
|                      | 14‡| US-Bi2: Bouldin Island | Apr 2017 – May 2018          |                               |
|                      | 15‡| D01: Union Island      | Apr 2016 – Sep. 2016         |                               |
|                      | 16‡| D06: Holland Island    | Jul. – Oct. 2016             | Medellín-Azuara, J. et al., 2018 |
|                      | 17‡| D08: Bouldin Island    | Jul. – Sep. 2016             |                               |
|                      | 18‡| D09: Bacon Island      | Jul. – Oct. 2016             |                               |
|                      | 19‡| D11: Staten Island     | Apr. – Sep. 2016             |                               |
| Pasture (697)        | 20‡| D03: Staten Island     | May 2016 – May 2017          |                               |
|                      | 21‡| D04: Jersey Island     | Jun. – Nov. 2016             | Medellín-Azuara, J. et al., 2018 |
|                      | 22‡| D05: Ripon Island      | Jun. 2016 – Jan. 2017        |                               |
|                      | 23‡| D12: Twitchell Island  | Jun. 2016 – Dec. 2016        |                               |
| Rice (2909)          | 24*| US-Twt: Twitchell Island| Apr. 2009 – Apr. 2017        | Baldocchi et al., 2016        |
| Tomato (112)         | 25‡| RR2: Russell Ranch     | Jan. 2017 – Nov. 2017        | Cawse-Nicholson et al., 2017. |
| Winter wheat (102)   | 26‡| RR1: Russell Ranch     | Jan. 2017 – Oct. 2017        |                               |

*Eddy covariance. †Full surface renewal system. ‡Lite surface renewal system estimates daily evapotranspiration by assuming the daily mean G is zero. †Reserved for independent validation.

Table 2. Comparison of daily evapotranspiration estimates from three Priestley Taylor approaches with field measurements during days with clear-sky Landsat observations and other days (from temporal interpolation). For clear-sky overpassing days, statistics were based on the full dataset, the testing dataset only (in parenthesis), respectively.

| Crop types | R² | RMSE (mm day⁻¹) |
|------------|----|-----------------|
|            | PT-0 | PT-JPL | PT-UCD | PT-0 | PT-JPL | PT-UCD |
|            | Clear Sky Overpassing Days | Interpolated | Clear Sky Overpassing Days | Interpolated | Clear Sky Overpassing Days | Interpolated |
| Alfalfa    | 0.64 (0.57) | 0.69 (0.71) | 0.83 (0.82) | 0.75 | 0.94 (1.05) | 0.87 (0.86) | 0.65 (0.68) | 0.77 |
| Almond     | 0.78 (0.78) | 0.48 (0.53) | 0.85 (0.86) | 0.89 | 1.13 (1.18) | 1.73 (1.72) | 0.95 (0.94) | 0.86 |
| Citrus     | 0.40 (0.24) | 0.60 (0.56) | 0.70 (0.70) | 0.70 | 0.99 (1.08) | 1.39 (1.48) | 0.69 (0.83) | 0.86 |
| Corn       | 0.18 (0.11) | 0.72 (0.70) | 0.75 (0.71) | 0.68 | 1.65 (1.66) | 0.97 (0.97) | 0.91 (0.96) | 1.01 |
| Pasture    | 0.75 (0.66) | 0.42 (0.36) | 0.77 (0.72) | 0.69 | 0.86 (0.86) | 1.31 (1.81) | 0.82 (0.77) | 0.94 |
|                | Rice      |          |          |        |        |        |        |
|----------------|-----------|----------|----------|--------|--------|--------|--------|
| All (Crop Specific) | N/A    | N/A      | 0.81 (0.79) | 0.80   | N/A    | N/A    | 0.87 (0.90) | 0.93 |
| All (Generalized)   | 0.70 (0.57) | 0.70 (0.71) | 0.80 (0.76) | 0.78   | 1.15 (1.32) | 1.12 (1.07) | 0.91 (0.98) | 0.95 |

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