Purification and Physical–Chemical Characterization of Bromocresol Purple for Carbon System Measurements in Freshwaters, Estuaries, and Oceans

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ABSTRACT: This work provides an algorithm to describe the salinity (S) and temperature (T) dependence of the equilibrium and molar absorptivity characteristics of purified bromocresol purple (BCP, a pH indicator) over a river-to-sea range of salinity (0 ≤ S ≤ 40). Based on the data obtained in this study, the pH of water samples can be calculated on the seawater pH scale as follows: pHSW = −log(Kc1 + log((R − e1)/(1 − e1))) where −log(Kc1) = 4.981 − 0.1710S + 0.09428S2 + 0.3794S1.5 + 0.00091229S2 + 310.2/T − 17.33S1/3/T − 0.5859S1/10/T. The term pHSW is the negative log of the hydrogen ion concentration determined on the seawater pH scale; R is the ratio of BCP absorbances (A) at 432 and 589 nm; Kc1 is the equilibrium constant for the second BCP dissociation step; and e1, e2, and e4 are BCP molar absorptivity ratios. A log(Kc1e1) equation is also presented on the total pH scale. The e1 value determined for purified BCP in this study can be used with previously published procedures to correct BCP absorbance measurements obtained using off-the-shelf (unpurified) BCP. This work provides a method for purifying BCP, fills a critical gap in the suite of available purified sulfonephthalein indicators, enables high-quality spectrophotometric measurements of total alkalinity, and facilitates pH measurements in freshwater, estuarine, and ocean environments within the range 4.0 ≤ pH ≤ 7.5.

INTRODUCTION

Sulfonephthalein pH indicators have been extensively used to describe the acid–base chemistry of oceans, estuaries, and rivers.1–5 Meta-cresol purple (mCP), for example, with physical–chemical characteristics particularly suitable for measurements at circumneutral pH (Table 1), has been widely used to obtain full water-column pH profiles in all five ocean basins and also to monitor pH in estuarine, freshwater, and sub-zero environments.6–13 Thymol blue (TB), with a pH-indicating range approximately 0.5 pH units higher than mCP,5,14 has been used to measure pH in cold open-ocean surface waters of New Zealand, the Norwegian Coastal Current, and the Weddell Sea.15–17 TB has also been employed in coastal systems with high photosynthetic activity (resulting in elevated pH) and in highly alkaline environments such as tidal pools of the San Juan Islands.7,18 Sulfonephthalein indicators with pH-indicating ranges lower than that of mCP, such as cresol red (CR) and phenol red (PR), have been used to measure freshwater pH.1,9,10 Additionally, CR has been used to study carbon chemistry dynamics under sea ice at high latitudes,20,21 and also pH distributions resulting from hydrothermal inputs on the Juan de Fuca ridge.22 Sulfonephthalein pH indicators have also been used for accurate and precise determination of other carbon system parameters such as total alkalinity (A),22,29 total dissolved inorganic carbon (CT),30,31 carbon dioxide fugacity (fCO2),32–34 seawater calcium carbonate saturation states (Ω)29, and the organic alkalinity of coastal seawater.35 Sulfonephthalein indicators have been used in some cases for in situ measurements, providing carbon system measurements with high spatial and temporal resolution.32,36–41 The diversity of uses for sulfonephthalein indicators has included, as well, investigations of acid–base equilibria and trace metal speciation42,43, observations of the hydration and dehydration kinetics of aqueous CO2,44,45 assessments of boron isotopic equilibria for determining paleo-pH,45,46 analysis of CO2-concentrating mechanisms in biota,47,48 and examinations of acid–base chemistry and metal toxicity in soils.49,50

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remove the substantial colored-impurities in some batches of commercial indicators that lead to erroneous absorbance ratio measurements, in conjunction with detailed characterizations of the physical—chemical properties of the indicator. Over the past 30 years, researchers have created algorithms to describe the behavior of sulfonephthalein indicators as a function of practical salinity ($S_p$) and temperature ($T$) (Table 1). Purification methods have been developed for four sulfonephthalein indicators (mCP, TB, PR, and CR), but only TB and mCP have been characterized over a freshwater-to-seawater range of salinities ($0 \leq S_p \leq 40$). The characteristics of purified PR have been reported only for zero ionic strength ($I$), and those for purified CR have been reported only for $S_p > 20$.

Table 1 lists the currently available suite of sulfonephthalein indicators in order of dissociation constant ($K_a$) expressed as $pK_a$ ($i.e., -\log K_a$) at 298.15 K. As a general guide, the pH-indicating range of each dye extends from approximately one pH unit above to one pH unit below its $pK_a$ value. $^{22,25,26,51}$ Significantly, the $K_a$ values of the first four indicators (TB, mCP, CR, and PR) range over only one log unit, while the difference between the values of the next two indicators [PR and bromocresol purple (BCP)] is nearly two units. As such, there is a substantial gap in the sulfonephthalein toolbox between purified indicators appropriate for mildly alkaline conditions ($7.5 \leq pK_a \leq 8.5$; TB, mCP, CR, and PR) and those appropriate for mildly acidic conditions ($4.0 \leq pH \leq 6.3$; BCP, bromocresol green (BCG)). Importantly, BCP is one of only two sulfonephthalein indicators with $pK_a$ values low enough for measurements of residual acid in total alkalinity titrations. Although BCG is, like BCP, appropriate for use in alkalinity titrations, BCP has a significantly higher $pK_a$ value than BCG (Table 1), making it the ideal indicator for quantifying residual acid at a relatively high pH, thereby minimizing uncertainties in residual acid determinations. As such, although it would be useful to expand characterizations of purified PR to include marine and estuarine conditions, and CR to include $S_p < 20$, the most pressing need in terms of current measurement capabilities is the development of BCP purification procedures and characterization of purified BCP. This new capability would extend the use of BCP for pH and $A_T$ measurements to include estuarine environments and enable pH measurements in environments that cannot be accessed with any of the current suites of purified sulfonephthalein indicators (e.g., alpine lakes or waters exposed to acid mine drainage).

| indicator | references | conditions: $S_p$, $T$ (K) | $\lambda$ (nm) | $pK_a$ ($S_p = 35$, $T = 298.15$ K) | pH ($R = 1$, $S_p = 35$, $T = 298.15$ K) |
|-----------|------------|-----------------------------|-----------------|--------------------------------|----------------------------------|
| TB        | Zhang and Byrne (1996) $^{14}$ | $30 \leq S_p \leq 40, 278.15 \leq T \leq 308.15$ | 435, 596 | 8.5 | 8.2 |
|           | Mosley et al. (2004) $^7$ | $0.00 \leq S_p \leq 40, T = 298.15$ | | | |
|           | Hudson-Heck and Byrne (2019) $^{2,3}$ | $0 \leq S_p \leq 40, 278.15 \leq T \leq 308.15$ | | | |
| mCP       | Clayton and Byrne (1993) $^{25}$ | $30 \leq S_p \leq 37, 293 \leq T \leq 303$ | 434, 578 | 8.0 | 7.6 |
|           | Mosley et al. (2004) $^7$ | $0.00 \leq S_p \leq 40, T = 298.15$ | | | |
|           | Lin et al. (2011) $^{24,44}$ | $20 \leq S_p \leq 40, 278.15 \leq T \leq 308.15$ | | | |
|           | Lai et al. (2016) $^9$ | $S_p = 0, 278.15 \leq T \leq 308.15$ | | | |
|           | Loucaides et al. (2017) $^3$ | $35 \leq S_p \leq 100$, freeing point $\leq T \leq 298.15$ | | | |
|           | Douglas and Byrne (2017b) $^3$ | $0 \leq S_p \leq 40, 278.15 \leq T \leq 308.15$ | | | |
|           | Müller and Rehder (2018) $^4$ | $0 \leq S_p \leq 40, 278.15 \leq T \leq 308.15$ | | | |
| CR        | Byrne and Breland (1989) $^{22}$ | $S_p = 35$, $T = 298.15$ | 433, 573 | 7.8 | 7.4 |
|           | French et al. (2002) $^{19}$ | $S_p = 0$, $T = 293.15$ | | | |
|           | Patsavas et al. (2013b) $^{41}$ | $20 \leq S_p \leq 40, 278.15 \leq T \leq 308.15$ | | | |
| PR        | Robert-Baldo et al. (1985) $^{22}$ | $33 \leq S_p \leq 37, 273 \leq T \leq 303$ | 433, 558 | 7.5 | 7.0 |
|           | Lai et al. (2016) $^9$ | $S_p = 0$, $281.15 \leq T \leq 303.15$ | | | |
|           | Yao and Byrne (2001) $^{26}$ | $S_p = 0$, $281.15 \leq T \leq 303.15$ | | | |
| BCP       | Breland and Byrne (1992) $^{27}$ | $29 \leq S_p \leq 35.2, 286.15 \leq T \leq 305.15$ | 432, 589 | 5.8 | 5.4 |
|           | Yao and Byrne (2001) $^{26}$ | $S_p = 0$, $283.15 \leq T \leq 303.15$ | | | |
|           | this work $^8$ | $0 \leq S_p \leq 40, 278.15 \leq T \leq 308.15$ | | | |
| BCG       | Breland and Byrne (1993) $^{28}$ | $29 \leq S_p \leq 37, 286 \leq T \leq 305$ | 444, 616 | 4.3 | 3.9 |

“Publications listed in bold font used purified indicator dye; asterisks denote publications that describe purification procedures. Publications that provide characterizations appropriate for freshwater [e.g., Lai et al. (2016) $^9$] will have higher corresponding $pK_a$ values. pH values were calculated using the bolded references for each dye with the exception of the pH value for PR, which was calculated utilizing Robert-Baldo et al. (1985).$^{25}$

### THEORY

The pH of aqueous solutions can be calculated from sulfonephthalein absorbance ratios ($R$) $^{1,4,5,14,24}$

$$pH = -\log(K_a c_2) + \log((R - e) / (1 - R e))$$

where $pH$ is the negative log of the hydrogen ion concentration ($-\log[H^+]$), expressed on either the seawater scale ($pH_{SW}$), de

or the total scale (pH$_T$), defined as [H$^+$]$_T$ = [H$^+$]$_S$ + [HSO$_4^-$]$_T$; K$_S$ is the equilibrium constant for the second dissociation step of the indicator dye, expressed on the seawater scale ($K_{2w}$) or the total scale ($K_2$), with units of mol/kg. R is a sulfonaphthalein absorbance ratio that is measured, in the case of BCP, at 432 and 589 nm (R = 589/A$_{432}$/432). K$_S$ are molar absorptivity coefficients, also expressed in terms of the absorbance properties of BCP at 432 and 589 nm. For BCP, these constants are defined as

\[
e_1 = \frac{589 e_{432} - 432 e_{589}}{\text{nm}} \quad (2a)
\]

\[
e_2 = \frac{589 e_{589} - 432 e_{432}}{\text{nm}} \quad (2b)
\]

\[
e_3 = \frac{432 e_{432} - 589 e_{589}}{\text{nm}} \quad (2c)
\]

\[
e_4 = e_3/e_2 = \frac{432 e_{432} - 589 e_{589}}{589} \quad (2d)
\]

where epsilon ($e$) has units of kg per mole per cm and the ratios are dimensionless.

Values of $e_1$ (by definition, dependent on the HI$^-$ species alone) can be determined under acidic conditions, where absorbance contributions from H$_2$I and I$^-$ are negligible. Similarly, values of $e_4$ can be determined at a sufficiently high pH that absorbance contributions from H$_2$I and HI$^-$ are negligible. Use of eq 1 obviates the necessity for direct determinations of $e_2$ and $e_3$ and thus reduces the number of required $e$ characterizations for spectrophotometric pH measurements from 3 to 2.

The log($K_{2e}$) term in eq 1 can be determined spectrophotometrically via paired measurements of mCP and BCP absorbance ratios (mCP/R and BCP/R). This type of approach is possible because there is a small region of overlap in the pH-indicating range of mCP and BCP. First, for a given batch of sample seawater and set of ($S_p$, $T$) conditions, the pH term in eq 1 is directly measured using mCP, with the absorbance ratio mCP/R = $A_{578}/A_{432}$ serving as input to the pH algorithm of Müller and Rehder (2018) (their eq 6 and Table 2; mCP/R).

### Table 2. Mobile Phase Profile for Purifying BCP Using a Redisep Gold C18Aq Column

| time (min) | % ACN |
|-----------|-------|
| 0–3       | 10    |
| 3–7       | 15    |
| 7–10      | 20    |
| 10–14     | 30    |
| 14–18     | 40    |
| 18–22     | 80    |
| 22–25     | 10    |

Then, for the same conditions (i.e., another sample of the same seawater at the same $S_p$ and $T$), the BCP absorbance ratio (BCP/R = 589/A$_{432}$) is measured. Finally, with known values of BCP $e_2$ (eqs 2a and 2d), eq 1 can be solved for log($K_{2e}$). It should be noted that log($K_{2e}$) is determined as a single entity in order to eliminate the need for independent determinations of $K_2$ and $e_2$.

### RESULTS

**Purification of BCP.** Table S1 outlines the purification method, using a Sielc PrimeSep B column, that provided the purified BCP used for the characterization of the absorbance and equilibrium properties of the indicator. During purification trials, it was noted that BCP had a very high affinity to the Sielc PrimeSep B column, resulting in a portion of the dye being inextricably retained on the column. As a result, the methods given in Table S1, although effective in purifying BCP, produced weight percentage recoveries of purified BCP somewhat smaller than 1%. Accordingly, additional purification methods were explored and resulted in weight percentage recoveries on the order of 2%. This yield is sufficient for approximately 8000 pH measurements. Table 2 shows the optimized mobile phase profile used to purify BCP with a Redisep Gold C18Aq column. The mixture is composed of acetonitrile (ACN), Milli-Q water, and 0.5% trifluoro acetic acid (TFA), with ACN being increased throughout the purification run. The main dye band began to move down the column when the ACN was 20% or greater (Figure S1). As the main band (Figure S1, orange) reached the end of the column, the initial portion of the band (Figure S1, yellow) was collected (approximately 30 mL) and HPLC analysis demonstrated that there was a minor impurity in this portion of the band. This portion of the band was discarded. Collection of the pure indicator was initiated when the absorbance reached a maximum and continued until the absorbance fell to 90% of the maximum (collected volume approximately 120 mL). Chromatographs of BCP before and after purification using the Redisep Gold C18Aq column (Figure 1) indicate the success of the method in removing impurities from commercial BCP. Impurity peaks seen in the chromatograph of off-the-shelf BCP at approximately 23 and 28 min, which show absorbance near 400 nm (Figure 1a), are absent in the post-purification chromatograph (Figure 1b).  

**Molar Absorptivity Characteristics of BCP.** For $e_1$, the average value for 288 ≤ $T$ ≤ 305 K can be expressed as

\[
e_1 = 0.00049 \pm 0.00029 \quad (3)
\]

The full $e_1$ data set is provided in Table S2. For the temperature range 288 ≤ $T$ ≤ 305 K, $e_1$ values with temperature could not be discerned (Table S2). Given the very small value of $e_1$ and considering that previous studies have noted the small influence of salinity relative to temperature on molar absorptivity ratios, variations with salinity were not explored.

The dependence of $e_4$ on $S_p$ and $T$ over 5 ≤ $S_p$ ≤ 40 and 278.15 ≤ $T$ ≤ 308.15 K (Table S3, Figure S3) are well described by the following model

\[
e_4 = -7.101 \times 10^{-3} + 7.674 \times 10^{-5} S_p + 1.361 \times 10^{-7} S_p^2 \quad (4)
\]

The residuals from this fit are shown in Figure 2 as a function of $T$. Overall, 95% of the residuals (Figure 2) are within ±0.00035. Figure S3 shows the separate influences of $T$ and $S_p$ on $e_4$. Though values of $e_4$ were not determined for $S_p < 5$, the weak dependence of $e_4$ on $S_p$ allows satisfactory extrapolation to lower salinities.

**Equilibrium Characteristics of BCP.** The experimentally determined log($K_{2e}$) values (Table S4) were fit using the following equation

\[
-\log(K_{2e}) = A + BS^{0.5} + CS + DS^{1.5} + ES^2 + F/T + GS^{1.5}/T + HS^{1.5} \ln T + IS^{0.5} T \quad (5)
\]

The coefficient values generated from these fits are provided in Table 3 for both the total and seawater pH scales. Overall, 99% of the residuals (i.e., empirical log($K_{2e}$)—predicted
log($K_2^{SW}$) and log($K_2^T$) are within ±0.006 over the full range of $S_P$ and $T$. This range of residuals is consistent with previous indicator characterizations performed using TRIS buffers.4 Furthermore, good experimental control of solution temperature (with $T$ between the paired mCP and BCP measurements differing by only 0.02 K on average) minimized the pH error attributable to $T$ fluctuations to within ±0.0001 for a given $S_P$ and $T$.

Table 4 shows calculated log($K_2^{SW}$) and log($K_2^T$) values for freshwater and typical seawater at 298.15 K using Table 3 coefficients. These values serve as check values to ensure that eq 5 coefficients, and all other coefficients, are correctly entered into investigators’ computational programs.

**DISCUSSION**

**Comparison with Previous Studies. Molar Absorptivity ($e_1$).** Previous investigators used off-the-shelf BCP to determine $e_1$, $e_2$, and $e_3$ for a single set of conditions each: $S_P = 35$, $T = 298.15$ K and $S_P = 0$, $T = 298.15$ K.26 This work, in contrast, used purified BCP to determine these values over ranges of $S_P$ and $T$ conditions.

For $e_1$, the value determined in this study (0.00049 ± 0.00029; eq 3) is roughly one-tenth the values reported by Breland and Byrne (1992)27 and Yao and Byrne (2001).26 We hypothesize that this difference is due to insufficient acidification in those earlier studies. The previous $e_1$ determinations were made in solutions in which the absorbance of HI$^-$ ($\lambda = 432$ nm) was maximized. However, further acidification is required to reduce absorbance contributions of the $F^-$ species at 589 nm to zero. The significance of this problem was not recognized in the previous work. Our investigations revealed that, subsequent to acidification to a point that HI$^-$ reached a maximum, further acidification to pH values less than 2,
the absorbance of HI− at λ = 432 nm was very nearly constant (i.e., H2I was not significantly impacting the measurements), while absorbances at λ = 589 nm decreased substantially (Tables S2 and S5). At pH ≈ 1.6, [I2−]/[HI−] ≈ 10−4, which resulted in ε1 values barely distinguishable from zero. The very low values obtained for ε1 in this work means that errors in ε1 will have a significant impact on pH calculations only at very low pH conditions (i.e., very low R values). As an example, the absorbance of HI− at λ = 432 nm was very nearly constant (i.e., H2I was not significantly impacting the measurements), while absorbances at λ = 589 nm decreased substantially (Tables S2 and S5). At pH ≈ 1.6, [I2−]/[HI−] ≈ 10−4, which resulted in ε1 values barely distinguishable from zero. The very low values obtained for ε1 in this work means that errors in ε1 will have a significant impact on pH calculations only at very low pH conditions (i.e., very low R values). As an example, the

Table 4. Check Values for Seawater and Freshwater Conditions

| S/P | T     | ε1   | ε2   | −log(K^[SW]2) | −log(K^[T]2) |
|-----|-------|------|------|---------------|--------------|
| 35  | 298.15| 0.00049| 0.0163| 5.3944        | 5.3850       |
| 0   | 298.15| 0.00049| 0.0158| 6.0214        | 6.0211       |

*Approximate pK2 values for BCP (5.85, at S/P = 35 and T = 298.15 K) can be obtained using the ε2 value of Breland and Byrne (1992).*

Figure 3. Residuals from fitting eq 5 to the log(K^[SW]2) data set (Table S4) shown here as a function of S/P.

Figure 4. Dependence of pH on S/P and T, as calculated using the BCP algorithm of this study (eqs 1, 3, 4, and 5) and the minimum and maximum R values that can be reliably measured with a typical seagoing spectrophotometer. The upper panels (a,b) are for seawater, and the lower panels (c,d) are for freshwater. The left panels (a,c) are for the case of R = 0.05, and the right panels (b,d) are for R = 20.

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difference in pH calculated using $e_i = 0.00049$ compared to $e_i$ values 30% larger and 30% smaller is $<0.00001$ at pH 6 and increases to only 0.001 at pH 4.

For $e_i$, the value experimentally determined in this work at $S_p = 35$ and $T = 298.15$ K ($e_i = 0.0162$; eq 4) is approximately 10% lower than the value calculated by Breland and Byrne (1992)\textsuperscript{27} ($e_i = 0.0178$) likely due to the absence of dye impurities in this work (Figure 1). Impurities in sulfonephthalein indicators characteristically absorb at short wavelengths (e.g., $\lambda = 432$ nm; Yao et al., 2007\textsuperscript{55} Figure 1). Given the very low absorbance of $I^−$ at $\lambda = 432$ nm (i.e., $A_{432} = 0.016$ when $A_{589} = 1.000$), any light-absorbing impurities will cause erroneously high absorbances at 432 nm and thereby erroneously high values of $e_i$.

$\log(K_{SW}\epsilon_i)$ Calculations. Values of $\log(K_{SW}\epsilon_i)$ for Breland and Byrne (1992)\textsuperscript{27} and Yao and Byrne (2001)\textsuperscript{56} were derived from their previous separate determinations of $\epsilon_i$ and $pK_e$. Figure S4 shows that the $-\log(K_{SW}\epsilon_i)$ values predicted from eq S5 for $T = 298.15$ K are, on average, 0.032 higher than the results of Breland and Byrne (1992)\textsuperscript{27} (comparing over the range $29 \leq S_p \leq 35$) and 0.016 lower than the $I = 0$ value of Yao and Byrne (2001).\textsuperscript{51} In view of the substantial methodological differences among the three studies, this level of agreement is remarkably good. Yao and Byrne (2001),\textsuperscript{51} for example, used phosphate buffer characteristics appropriate at low ionic strength to calculate their $\log(K_{SW}\epsilon_i)$ values, whereas the BCP $\log(K_{SW}\epsilon_i)$ values of this work (eq S5, Table 3) are directly dependent on mCP $\log(K_{SW}\epsilon_i)$ indicator properties that were determined using TRIS buffers characterized with Harned cells.\textsuperscript{4}

It is important to note that the uncertainty in BCP $\log(K_{SW}\epsilon_i)$ values is directly linked to the uncertainty in $m_{CP}\log(PH)$ (approximately 0.010; Orr et al. 2018).\textsuperscript{53} As such, the uncertainty of pH values obtained with BCP will be somewhat greater than 0.01 pH units. Notably, because the algorithms provided in this study (eqs 3, 4, and 5) are based on the molecular properties of purified BCP and are linked to the characterization of mCP, if BCP coefficients (Table 3) or mCP coefficients are refined in the future, historical pH data obtained using purified BCP can accordingly be easily revised.

Use of BCP to Measure pH. The pH-indicating range of any particular pH indicator dye depends on how $\log(K_{SW}\epsilon_i)$ varies with $T$ and $S_p$. The algorithm developed in this work describes, for the first time, BCP $\log(K_{SW}\epsilon_i)$ values over $0 \leq S_p \leq 40$ and $278.15 \leq T \leq 308.15$ K, thereby providing pH measurement capabilities over wide-ranging conditions in rivers, estuaries, and oceans. Notably, this work provides, also for the first time, a basis for spectrophotometric determinations of alkalinity that include estuarine conditions, as well as procedures for eliminating alkalinity errors associated with the use of impure BCP.

The mid-point of the pH-indicating range of a sulfonephthalein indicator dye is determined by the physicochemical properties of the dye—specifically, its $pK_e$ value (Table 1) as well as $e_i$. The extent of the pH-indicating range (narrow or wide) about that mid-point is determined by the quality of the spectrophotometer used to measure $A$ for a given application or experiment\textsuperscript{54} and is also influenced by the $e_i$ value of the dye. For the purpose of characterizing indicator dyes (such as methods utilized in this work), it is essential that high-quality spectrophotometers (i.e., allowing measurements at $A > 3$) are used. However, accurate measurements of pH can still be achieved for general purposes using moderately priced spectrophotometers. If measurements are obtained using a lower specification spectrophotometer capable, nonetheless, of accurately measuring absorbances over an absorbance range of 0.05 $\leq A \leq 1$ (e.g., the Agilent 8453, which is often used for shipboard pH measurements), then corresponding conservative assessments of BCP absorbance ratios (0.05 $\leq R \leq 20$) can be used in conjunction with known BCP properties (eqs 3, 4, and 5 and Table 3) to describe the BCP pH-indicating range as a function of $S_p$ and $T$ (Figure 4). For seawater of $S_p = 35$ and $T = 298.15$ K, the pH-indicating range of BCP is thus shown to extend from pH 4.08 (Figure 4A) to 6.86 (Figure 4B). For freshwater ($S_p = 0$, $T = 298.15$ K), the pH-indicating range extends from pH 4.72 (Figure 4C) to 7.49 (Figure 4D), on the order of 0.6 units higher than for seawater due to changes in indicator and $H^+$ activity coefficient characteristics between low and high salinity waters.

As discussed in Hudson-Heck and Byrne (2019),\textsuperscript{8} a more expensive and higher-quality spectrophotometer (e.g., the Cary 400, used for the benchtop studies of this work with periodically verified linearity) can enable accurate measurements at higher absorbances (e.g., $A > 3$) and therefore extend the range of $R$ values used in pH measurements. In this case, the BCP pH-indicating range would be expanded substantially beyond what is shown in Figure 4. However, Hudson-Heck and Byrne (2019)\textsuperscript{8} also noted that the denominator of eq 1 indicates that $R$ cannot exceed $e_i^{-1}$ (i.e., $R_{bc}$ must be $\leq 1$) and, as such, indicator dyes have an inherent maximum $R$ value that is directly dependent on the magnitude of the $e_i$ value of that dye. Because of this limitation, inaccuracies in pH calculations can become large as $R$ closely approaches its maximum value.

Measurements of both mCP and BCP absorbance ratios performed in this study were made in solutions with $6.2 \leq pH \leq 7.0$. Although measurements of $m_{CP}R$ were made at a pH slightly lower than the ideal indicating range of mCP, the magnitude of $m_{CP}R_{bc}$ in the denominator of eq 1 was small which minimized the uncertainty in $m_{CP}\log(PH)$ calculations. Conversely, measurements of $bc_{CP}R$ were made at a pH slightly higher than the ideal indicating range of BCP. However, the BCP $e_i$ value is uniquely low ($<0.005$) and therefore BCP has a much higher inherent maximum $R$ value than most other indicators. Therefore, since measurements of mCP and BCP were performed within appropriate ranges of $R$ (determined by the value of $e_i$ and the quality of the spectrophotometer), the upper bound uncertainty of the log $R$ term in eq 1 [i.e., $\log((R - e_i)/(1 - R_{bc}))$] can be estimated as $\pm 0.005$. In this case, combining the $\pm 0.01$ uncertainty of mCP pH\textsuperscript{53} with $\pm 0.005$ uncertainties for the log($((R - e_i)/(1 - R_{bc}))$ terms of both mCP and BCP, the uncertainty of the BCP $\log(K_{SW}\epsilon_i)$ is calculated as $\pm 0.012$.

In this work, we measured $e_i$ values of two unpurified batches of BCP (Kodak batch A8a and TCI batch WU III-FQ). Both batches of unpurified BCP showed very high levels of impurities compared to the pure $e_i$ value determine in this study (pure = 0.016, Kodak = 0.133, and TCI = 0.165). Subsequently, these two unpurified batches of BCP were used to measure spectrophotometric pH along with corresponding pH measurements made with pure BCP. These measurements, performed in 0.7 M NaCl solutions over a range of pH from 3.8 to 5.3, highlight the large pH errors that can arise through the use of unpurified commercial batches of BCP. The use of unpurified BCP for these two batches of indicators (Kodak and TCI) produced pH errors (differences between pure and impure indicator) as large as 0.07 at pH 3.8 and as large as 0.2
at pH 5.3. These errors in measured pH using batches of unpurified commercial BCP are much greater than what has been observed for unpurified mCP.\textsuperscript{24,55} Notably, however, as the BCP characterizations of Yao and Byrne (2001)\textsuperscript{26} and Breland and Byrne (1992)\textsuperscript{27} are in substantial agreement with the purified-BCP characterizations obtained in the present work, the characteristics of commercial BCP are diverse and can include batches with low levels of impurities. Accordingly, the use of purified BCP is certainly the preferred option, and the use of unpurified batches of BCP should include purity-assessments. As a tool to quickly assess the impurity of a given batch of BCP, we suggest researchers perform measurements of $e_i$ and compare the results to $e_f$ values for pure BCP (eq 4). In addition, researchers should examine the absorbance spectra of each batch of BCP at pH $\sim$1.6 to confirm that the wavelength of maximum absorbance is at 432 nm.

For best practices, we recommend using purified BCP (not currently commercially available) to measure the pH of aqueous samples, but we also recognize that the process of purifying an indicator is laborious and time-consuming and may be out of reach for some investigators or even unnecessary for some applications. Douglas and Byrne (2017a)\textsuperscript{56} have outlined an alternative approach, whereby accurate pH measurements with mCP can be achieved via (a) absorbance ratios obtained using an off-the-shelf indicator in combination with (b) absorbance ratio corrections obtained using $e_i$ values appropriate to the purified form of that indicator. Our eq 4 can be used with the procedure outlined in Table 1 of Douglas and Byrne (2017a)\textsuperscript{56} to correct absorbance ratios obtained with unpurified BCP. This method is described in detail in Douglas and Byrne (2017a).\textsuperscript{56} As a brief summary, (1) solutions of 0.7 M NaCl at pH 12 are prepared, (2) $R$ values are measured using unpurified BCP, (3) absorbance contributions from impurities are calculated using eq 17 of Douglas and Byrne (2017a),\textsuperscript{56} and (4) using the absorbance contribution from the impurity at 432 nm, $R$ values obtained with unpurified BCP are corrected to $R$ values appropriate to purified BCP. It is important to note that this type of approach is appropriate for batches of unpurified BCP with moderate levels of impurities [such as those used by Breland and Byrne (1992)\textsuperscript{27} and Yao and Byrne (2001)].\textsuperscript{26} This correction procedure may be less effective for correcting BCP pH measurements that are obtained with high levels of impurities such as Kodak A8a and TCI WU III-FQ batches.

Use of BCP to Measure $A_T$. Single-step spectrophotometric $A_T$ measurements\textsuperscript{2,26} rely on interpretations of BCP absorbance ratios and subsequent calculations of pH in order to quantify the residual acid that remains in a sample after acidification and then bubbling to remove CO$_2$. To reduce errors in residual acid, we recommend using purified BCP. If, however, purified BCP is unavailable, then this work’s characterization of $e_i$ (eq 4) can be used in combination with the procedure of Douglas and Byrne (2017a)\textsuperscript{56} to reduce errors caused by the use of off-the-shelf BCP for determinations of $A_T$.

As an additional means of reducing $A_T$ errors in single-step acid addition methods,\textsuperscript{2,29} we recommend that titrations be performed such that the final pH (i.e., after acidification and bubbling) is above 4.5 (i.e., minimizing the concentration of the residual acid). In this case, errors in the measured pH [due to errors in log($K_e$)] propagate to produce only small errors in derived $A_T$ (Figure 5).\textsuperscript{2,55} Even for a systematic pH error as large as 0.02 (Figure 5 blue line), if the final pH is $>$4.5 then the contribution of this 0.02 unit pH error in the excess acid term to an error in $A_T$ is $\leq$1.8 $\mu$mol/kg. Accordingly, the use of impure indicators for single-step $A_T$ measurements, in conjunction with the correction procedure of Douglas and Byrne (2017a),\textsuperscript{56} should be sufficient to achieve accuracy well within the $\pm$2 $\mu$mol/kg uncertainties typical of modern $A_T$ analyses.

■ CONCLUSIONS

Purified and well-calibrated indicator dyes are essential, high-quality analytical tools for obtaining measurements of pH and other carbon system parameters in aqueous solutions. This study adds BCP to the suite of available sulfonephthalein indicators (Table 1) by (a) developing a method to purify BCP, (b) providing a characterization of purified BCP over 0 $\leq$ $S_p$ $\leq$ 40 and 278.15 $\leq$ T $\leq$ 308.15 K, and (c) reporting a key parameter ($e_i$) needed to make spectrophotometric absorbance measurements using off-the-shelf BCP. This work thus enables accurate BCP-based pH measurements in freshwater, estuarine, and marine conditions; expands the use of sulfonephthalein indicators to include pH measurements in mildly acidic environments (e.g., alpine lakes, soils, and waters impacted by acid-mine drainage); and improves the accuracy and range of conditions that can be utilized in spectrophotometric $A_T$ measurements.

■ EXPERIMENTAL SECTION

Materials and Reagents. BCP (acid form) (Tokyo Chemical Industry, TCI, Batch WU III-FQ), mCP sodium salt (TCI, Batch M0074), high-performance liquid chromatography (HPLC)-grade ACN, high-purity TFA, and ultrapure bis–tris (≥98%) were purchased from Fisher Scientific. High purity sodium chloride (NaCl), calcium chloride ($\text{CaCl}_2$), and potassium chloride (KCI) salts were purchased from MP Biomedicals, Sigma Aldrich, and Fisher Scientific, respectively. Seawater was collected from the surface waters of the open Gulf of Mexico. Purified mCP was obtained using the flash...
BCP purification was performed using a Teledyne ISCO CombiFlash RF instrument and purification quality was assessed using a Waters Prep HPLC system. Multiple columns with different solid-phase compositions (e.g., Sielc PrimeSep B, Redisep Gold C18Ag, and Redisep Gold C18) were explored to optimize the purification method. The SP of each seawater solution was measured (±0.01) on a Guildline 8410A salinometer, and the temperature was measured (±0.05) with a Fisher Scientific, Traceable thermometer. BCP and mCP absorbance measurements were conducted using a Cary 400 Bio UV-VIS dual-beam spectrophotometer (bandwidth = 0.1 nm). Glass spectrophotometric cells (10 cm pathlength) housed in a custom-made thermostatted cell holder inside the Cary 400 were equilibrated to the desired temperature using a Lauda Ecoline E-100 circulating water bath. Acid titrations (performed in the course of the εe and logKε2ε determinations) were monitored using an Orion pH electrode that had been spectrophotometrically calibrated using mCP.58,59

Absorbance Measurement Protocol. Each optical cell was first equilibrated to the desired T, and a blank absorbance measurement (i.e., solution only, with no added indicator dye) was recorded. For determinations of log(Kε2ε), paired cells of identical solution were used: one for the mCP measurements and one for the BCP measurements. After dye injection, absorbances were measured at the wavelengths of maximum absorbance for the dye (453λ and 579λ nm for mCP, and 432λ and 589λ nm for BCP). To account for pH perturbations resulting from indicator additions, each cell received two dye injections;23,24 observed R values could then be extrapolated to R values appropriate to zero added indicator. To correct for potential baseline shifts during measurements, absorbances at non-absorbing wavelengths (A434 for mCP and A432 for BCP) were also measured and used in the calculation of absorbance ratios

\[
\text{mCP}R = \frac{A_{579} - A_{730}}{A_{434} - A_{730}} \quad \text{and} \quad \text{BCP}R = \frac{A_{589} - A_{750}}{A_{432} - A_{750}}
\]

The volume of each indicator addition was chosen to maintain absorbance measurements within the linear range of the spectrophotometer (approximately 0.0–4.0 absorbance units for the Cary 400). For low-temperature conditions, dry nitrogen gas was directed at the cells’ optical surfaces to prevent condensation.

Purification of BCP. Multiple purification trials were performed to determine the optimal procedure for purifying BCP. Each purification began by dissolving unpurified indicator powder in Milli-Q water. The indicator was added to the purification column (~20 mL) as a stock solution of 50 mM BCP plus 0.5% TFA. There were no solubility problems at this concentration of BCP. This stock solution was then loaded onto a flash purification column that had been saturated with 5% ACN. Separation of pure product from the impurities was continuously monitored from a control screen, and once the purified dye eluted off the column, the eluate was collected and its purity was HPLC-verified (Sielc PrimeSep B2 column, 70%, ACN, 30% Milli-Q water, 0.1% TFA). Residual mobile-phase solvents were removed from the eluate through evaporation (i.e., air-dried), and the resulting purified solid was then redissolved in Milli-Q water (10 mM) to produce a pure indicator solution for use in the BCP characterization experiments. Dissolution of the purified BCP solid was facilitated by incremental additions of sodium hydroxide (NaOH; 1 M) until the dissolution was complete (typically 200–300 μL of NaOH was required).

Determination of εe. Measurements of εe (eq 2a) were conducted in a solution of NaCl (0.7 M; see the Results section for further details) using an open-top quartz spectrophotometric cell (10 cm pathlength) that was fitted with a lid to support an Orion pH electrode, an overhead stirring rod, and a temperature probe. BCP (~6 μM) was added to the NaCl solution. A strong acid (HCl, 1 N) was used to titrate the NaCl solution to a pH at which 589A/432A reached a minimum (usually 1 ≤ pH ≤ 2). At low pH, the solution is well buffered solely by the presence of H+. Values of εe were directly determined from absorbance ratios (εe = 589A/432A) measured over a range of temperature: 288 ≤ T ≤ 305 K. Anionic strength of 0.7 M is approximately equivalent to SP = 35 in seawater.

Determination of εe. Measurements of εe (eq 2d) were performed in artificial seawater at pH = 12, whereby [I2−]/[HI−] = 106. In a 4 L amber bottle, a stock solution of artificial seawater was prepared at SP = 40 (0.568 M NaCl, 0.072 M CaCl2, and 0.012 M KCl), largely following the recipe of DelValls and Dickson (1998).60 The MgCl2 and Na2SO4 of the original recipe were excluded from our formulation in order to avoid the formation of precipitates under the high-pH conditions required to maximize [I2−]. This stock artificial seawater was gravimetrically diluted to obtain a range of salinities (Table S4) and included the addition of sufficient NaOH (1 M) titrant to each solution to maintain a pH of 12 (0.01 M NaOH). Two dye injections were required for each εe measurement. After the first injection, 589A was measured. After the second injection, 432A was measured. The second addition of indicator was used to increase the BCP concentration by a factor of ~10 because 432εi− is very small relative to 589εi−. For both measurements, an additional wavelength (500λ) was monitored so that the ratio of absorbances after the second and the first injections (589A(2)/589A(1)) could be used to calculate the ratio of the BCP concentrations after the first and second indicator additions. This wavelength choice was not tightly constrained. Other wavelengths in the vicinity of 500 nm would have been equally suitable. Measurements were performed over ranges of salinity and temperature: 5 ≤ SP ≤ 40 and 278.15 ≤ T ≤ 308.15 K.

Determination of log(Kε2ε). Values of log(Kε2ε) were determined using paired absorbance measurements, as described in Hudson-Heck and Byrne (2019).5 A stock solution of CO2-free seawater was prepared by using HCl (1 N) to acidify the solution to pH = 4.2 and subsequently purging the solution of CO2 with dry nitrogen gas. This acid titration was monitored using an Orion pH electrode calibrated spectrophotometrically with mCP.58 Bis-tris was added to the CO2-free seawater (approximately 1 mM) and, using HCl (1 M) or NaOH (1 M), the pH was adjusted to an R ratio at which the absorbances of BCP and mCP were within the linear range of the spectrophotometer (typically, at 6.2 ≤ pH ≤ 7.0, it is observed that 0.05 ≤ mCP/R ≤ 0.07 and 7 ≤ BCP/R ≤ 11). Measurements of log(Kε2ε) at low ionic strengths (I ≤ 0.004) were performed in mixtures of Milli-Q water and bis-tris (1 mM). The ionic strengths and corresponding salinities of these solutions were calculated from the concentrations of added HCl.
The glass spectrophotometric cells (one cell for mCP and another cell for BCP) used for the measurements were filled with bis–tris buffered solutions (using Teflon tubing) and allowed to overflow for ~20 s. Absorbance measurements were collected following the protocol outlined in the Absorbance Measurement Protocol section. The paired measurements of mCP and BCP absorbance ratios were each obtained in triplicate. Each aliquot of sample solution received two dye injections, allowing extrapolation of observed absorbance ratios to R values appropriate to zero added indicator. The perturbation correction is essential for obtaining accurate absorbance ratios in weakly buffered solutions (i.e., seawater or Milli-Q water with bis–tris). The corrected R values for mCP and BCP were then used to calculate either the solution \( m_{\text{CP}} \) (mCP aliquot) or an absorbance ratio (BCP aliquot). Finally, using \( m_{\text{CP}} \) and BCP, values of log(\( K_{e2} \)) were calculated using eq 1 over ranges of salinity and temperature (0 ≤ S ≤ 40 and 278.15 ≤ T ≤ 308.15 K).

**Fitting Procedure for log(\( K_{e2} \)) Data Sets.** Using the programming code CO2SYS,\(^5\) \( K_{e2} \) results were reported on both the total and seawater scales for compatibility with previous results reported on these two pH scales: measurements of \( m_{\text{CP}} \) on the total scale were converted to equivalent values of \( m_{\text{CP}} \) on the seawater scale, and these values were subsequently used to determine log(\( K_{e2} \)) on the seawater scale. Both data sets (total and seawater scales) were fit as functions of salinity and temperature using the "stats" package of the programming language "R."\(^6\) Coefficients were added or subtracted from the model in a stepwise fashion until \( R^2 \) was minimized. For each set of triplicate R measurements, the average temperature over the three measurements was used.

**ASSOCIATED CONTENT**

1. **Supporting Information**
   The Supporting Information is available free of charge at https://pubs.acs.org/doi/10.1021/acsomega.1c01579.

2. **Supplemental purification method; data tables for \( e_1 \) and \( e_2 \) and log(\( K_{e2} \)) results; figure of various unpurified BCP batches; figure of flash column during purification; figure of \( e_1 \) model fits; and figure of log(\( K_{e2} \)) comparison data (PDF)**

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**Notes**
The authors declare no competing financial interest.

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