Naphthalene exerts substantial nontarget effects on soil nitrogen mineralization processes in a subalpine forest soil: A microcosm study

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Abstract

Naphthalene has been widely used to test the functional roles of soil fauna, but its nontarget effects remain uncertain in various soils. To determine whether there is a potential nontarget effect on soil biochemical properties in subalpine forest soil, soils in a subalpine forest on the western Qinghai-Tibet Plateau were treated by naphthalene in microcosms. The responses of soil microbial activity and nutrients to naphthalene were studied following 52 days of incubation. The results showed that the naphthalene application obviously decreased the microbial respiration rate in the first 10 days of the incubation and then increased the rate in the following days of the incubation. Moreover, the naphthalene application did not significantly affect the microbial activities overall, measured as soil microbial phospholipid fatty acid (PLFA) abundances and biomasses, or most enzyme activities (invertase, nitrate reductase and nitrite reductase) during the whole incubation period. However, naphthalene suppressed increases in the DON, NH4+-N and NO3--N contents and urease activity and led to the net mineralization of inorganic N (NH4+-N + NO3--N), in contrast to the net immobilization result in the controls. These results suggest that naphthalene can exert direct nontarget effects on soil microbial respiration and N mineralization processes in subalpine soils. Caution should be taken when using naphthalene to repel soil animals in field experiments.

Introduction

Soil biota (animals and microbes) are indispensable components and key drivers of soil biogeochemical processes, contributing significantly to organic matter decomposition, nutrient mineralization and greenhouse gas emissions in various ecosystems [1, 2]. While microorganisms are the primary regulators of most of these processes [2, 3], soil animals can play a
functional and direct or indirect role in biogeochemical cycling by fragmenting and comminuting litter, regulating microbial activities and altering soil aggregates and structure [4, 5]. Moreover, the composition, diversity and interactions of soil animals and microorganisms modulate carbon and nutrient cycling channels in soil detrital food chains [6, 7]. Consequently, carbon and nutrient cycling cannot be modeled precisely without a full understanding of the functional roles of faunas and their interactions with microorganisms in soil detrital food chains [8, 9].

In recent decades, various methods have been performed to characterize the functional roles of soil animals [10–12]. However, studying the functional roles of soil animals in soil biogeochemical cycling in situ without producing nontarget effects on other organisms or affecting the microclimate in surface soil remains challenging [3, 9]. Compared to other biocides (e.g., pyridaben, profenofos and triflumuron), naphthalene (C\textsubscript{10}H\textsubscript{8}) has long been considered to have fewer nontarget effects; this biocide has been used to repel soil fauna and to determine the functional roles of soil fauna in soil and litter organic matter decomposition processes in field experiments [12–14]. Nevertheless, it has long been suspected that naphthalene may indirectly influence these processes through its potential effects on soil microorganisms and nutrients [3, 15]. For example, early microcosm studies suggested that naphthalene might directly affect microbial populations and activities [14]. However, conflicting results showed that naphthalene treatment had weak direct influences on microbial phospholipid fatty acid (PLFA) abundances and carbon dynamics [3]. Moreover, the nontarget effects of naphthalene varied substantially with changes in soil types and incubation conditions [3, 15–17]. Scant research is available regarding the nontarget effects of naphthalene in high-latitude and high-altitude ecosystems, such as alpine and subalpine ecosystems. Consequently, there is a clear need for determining whether there is a direct nontarget effect on the soil biochemical properties in alpine and subalpine forests before conducting field experiments with naphthalene [14, 15].

Here, a microcosm experiment was conducted by adding naphthalene to subalpine forest soil in Southwest China, and the nontarget effects of such additions on soil microbial activities (respiration, PLFA abundance, microbial biomass and enzyme activity) and soil nutrients (dissolved organic matter and inorganic N) were determined. The aims were (1) to assess whether there was a potential nontarget effect on the soil biochemical properties in the subalpine forest soil and (2) to test the hypotheses that naphthalene application might have a stronger nontarget effect on the soil biochemical properties of nitrogen cycling processes than on those of carbon cycling processes because previous studies have suggested naphthalene has a stronger influence on soil inorganic nitrogen availability [14, 16].

**Materials and methods**

**Ethics statement**

We received permission from the Lixian Forestry Bureau to collect the tested soil in a local forest in 2015. In this study, only a limited number of soil samples were collected to conduct a microcosm experiment, and thus, our work had negligible influences on the function of the broader ecosystem. In addition, this study was carried out in compliance with the laws of the People’s Republic of China. The research did not involve measurements of humans or animals, and no endangered or protected plant species were involved.

**Experimental design**

In October 2015, approximately 20 kg of tested soil was collected in a secondary fir forest at the Long-term Research Station of Alpine Forest Ecosystems (31’18’N, 102’56’E, 3023 m a.s.l.), Southwest China. Soil was collected from five plots (2 m × 2 m size) in the forest using a soil
auger (15 cm depth and 5 cm diameter) and mixed after removing visible debris and fresh litter. The soil type was a Cambic Umbrisol according to the IUSS Working Group WRB [18], and the basic soil chemical properties (0–15 cm depth) were as follows: pH 6.5 ± 0.3, soil bulk density 1.04 ± 0.11 kg m⁻³, total organic carbon 153.9 ± 27.4 g kg⁻¹, total nitrogen 7.8 ± 1.3 g kg⁻¹ and phosphorus 0.9 ± 0.1 g kg⁻¹ [19]. The collected soil was sieved (2 mm) and then mixed. Stones, visible animal and plant residues and live macrofauna were removed (e.g. earthworms and millipedes), and the samples were air-dried for the soil microcosm experiment.

Nightly replicate microcosms were constructed with 450 ml clear glass culture bottles (17.5 cm high x 16 cm i.d.). Eighty of the microcosms were filled with 50 g of air-dried soil, and the remaining ten empty microcosms served as blanks. Microarthropods and nematodes in each soil microcosm were eliminated according to the description of Blair et al. [14]. In brief, the soil microcosms were microwaved in a 700-W microwave oven for 120 s and capped for 12 h. This microwave treatment was repeated three times. The soil moisture (w/w) in the microcosms was adjusted to 45% with a suspension of soil microorganisms. The suspension was prepared by homogenizing 200 g of sampled soil in 1.5 liters of deionized water and filtering the homogenate through a 5 μm nucleopore filter [14].

Soil incubation and respiration measurement

The microcosms consisted of a control group and a treatment group, with 45 microcosms (40 soil microcosms and 5 blanks) for each group. The treatment group then received 0.35 g of naphthalene per bottle at the beginning of the experiment. The other group served as a control (without naphthalene). The soil incubation lasted for 52 days, and two additional naphthalene applications (0.35 g) were performed on days 17 and 38. The total application rate is consistent with field application rates (100 g m⁻²) [14].

The soil microbial respiration rate was allowed to stabilize for three weeks. Before the incubation, plastic vials (8 cm high x 10 cm i.d.) containing 20 ml of 0.01 N NaOH were placed into culture bottles. Following the incubation, the culture bottles were capped and store at 10°C and 45% moisture in temperature-controlled biochemical incubators. The culture temperature and moisture were consistent with our previous field monitoring results [19]. The soil microbial respiration rate in each culture bottle was estimated by determining the carbon dioxide (CO₂) production following the sampling schedule for 52 days. CO₂ production was determined by titration with 0.02 N HCl following the addition of 1 ml of 1 N BaCl [20]. Empty bottles without soil were used as controls.

Soil sampling and chemical analysis

Soil sampling was performed on days 3, 10, 17, 24, 31, 38, 45 and 52 following the incubation. At each sampling time, 10 soil microcosms (5 controls and 5 naphthalene treatments) were sampled after soil respiration measurements were terminated. Fresh soils were kept in refrigerators at 4 ºC and -70 ºC for soil chemical and microbial analysis, respectively.

The soil extractable N (NH₄⁺-N and NO₃⁻-N) was extracted with 2 M KCl and then measured by the method of Lu [21]. The soil dissolved organic carbon (DOC) and dissolved total nitrogen (DTN) were extracted with 0.5 M K₂SO₄ [22, 23]. DOC and DTN in the extracts were quantified by a TOC–VcPH+TNM–1 C/N analyzer (Shimazu Inc., Kyoto, Japan). The soil dissolved organic nitrogen (DON) was calculated as follows: DON = DTN − NH₄⁺-N − NO₃⁻-N [23]. The soil microbial biomass carbon (MBC) and nitrogen (MBN) were determined by the chloroform fumigation extraction method with a conversion factor of 0.45 for MBC and 0.54 for MBN [24, 25]. The soil enzyme activities of invertase (mg glucose g⁻¹ soil dry weight (DW) d⁻¹) and urease (mg NH₄⁺-N g⁻¹ soil DW d⁻¹) were measured according to the methods of...
Wang et al. [26] and Lu [21], respectively. Soil nitrate reductase and nitrite reductase activities (mg NO$_2$·N g$^{-1}$ soil DW d$^{-1}$) were analyzed according to Xiong et al. [27].

The soil microbial PLFAs were extracted and quantified according to the methods of White et al. [28] and He et al. [29] with partial modifications. The sum of all subsequently described PLFAs was used as a proxy for the total microbial biomass. We used the sum of i15:0, a15:0, i16:0, i17:0 and a17:0 as gram-positive bacterial markers [30, 31], 16:1ω7c, 16:1ω9c, cy17:0, 18:1ω7c, and cy19:0 as gram-negative bacterial markers [32] and 15:0, 16:0, 16:1ω5t, 17:0, 18:00 and 20:5 as general bacterial markers [33, 34]. The gram-positive, gram-negative and general bacterial markers were summed to give the total bacteria. We used the sum of 18:3, 18:1ω9c, 18:2ω6, 9c and 20:1ω9c as fungal markers to represent the total fungi [35, 36]. The detailed gas chromatography–mass spectrometry (GC-MS) conditions were described by Liu et al. [9].

Data calculation and statistical analyses
The net ammonification, nitrification and inorganic N mineralization at the end of the incubation (52 days) was calculated as the differences in the NH$_4^+$-N (ammonium), nitrate (NO$_3^-$-N) and inorganic N (NH$_4^+$-N + NO$_3^-$-N) contents in the microcosms between the start and end of the incubation. Moreover, data for quantifying the effects of naphthalene addition on soil biochemical properties were calculated as the differences in the average values of the measured variables between the naphthalene treatments and the controls during the whole incubation period. For specific sampling times, Student’s independent-sample t-test was used to compare the effects of naphthalene application. We used repeated measures of analysis of variance (ANOVA) to test the effects of naphthalene application, sampling time, and their interactions on the measured variables. Differences were considered significant at $P < 0.05$ level for all analyses. All statistical analyses were performed using SPSS 18.0 software package for Windows (SPSS Inc., IL, USA).

**Results**

**Naphthalene effects on the soil microbial respiration rate**

The soil microbial respiration rates in both the naphthalene and control (without naphthalene) microcosms varied significantly ($F = 303.20$, $P = 0.001$) over time (Table 1). Tests for time dynamics showed a decreasing trend throughout the experiment (Fig 1). Compared to the control, the naphthalene application obviously ($F = 30.45$, $P = 0.001$) decreased the soil microbial respiration rates in the first 10 days following the incubation and then increased the rates to a higher level between days 24 and 52 of the experiment.

**Naphthalene effects on the soil microbial PLFAs**

The abundances of PLFAs (bacteria, fungi, gram-positive (G$^+$) bacteria and gram-negative (G$^-$) bacteria) were not significantly (all $P > 0.05$) affected by the naphthalene application, but the interaction effect of naphthalene application and sampling time was significant (all $P < 0.05$) on these PLFA abundances (Table 1). Compared to the control, the naphthalene application decreased the abundance of PLFAs and the ratio of fungal PLFAs to bacterial PLFAs before the third naphthalene application on day 38 (Fig 2A–2E). Then, the abundance and ratio of fungal PLFAs and bacterial PLFAs significantly increased until the end of the incubation (Fig 2A, 2B and 2E). Conversely, the ratio of G$^+$ to G$^-$ bacteria in the naphthalene microcosms increased before the third naphthalene application on day 38 but did not significantly differ in the following incubation (Fig 2F).
Table 1. Repeated measures ANOVA results of the responses of the soil respiration rate, microbial phospholipid fatty acid (PLFA) abundance, microbial biomass, nutrient content and enzyme activities to naphthalene application and sampling time.

| Variables                  | Naphthalene (N) | Time (T) | N × T |
|----------------------------|-----------------|----------|-------|
|                            | df  | F    | P    | df  | F    | P    | df  | F    | P    |
| Soil respiration rate      | 1   | 30.45 | <0.001** | 7   | 303.20 | <0.001** | 7   | 62.86 | <0.001** |
| Bacterial PLFAs            | 1   | 4.58  | 0.099 | 7   | 20.68  | <0.001** | 7   | 4.59  | 0.002** |
| Fungal PLFAs               | 1   | 0.36  | 0.593 | 7   | 9.38   | <0.001** | 7   | 5.92  | 0.001** |
| Fungal/bacterial PLFAs     | 1   | 0.36  | 0.570 | 7   | 2.35   | 0.040*   | 7   | 2.78  | 0.018*   |
| G+ PLFAs                   | 1   | 6.65  | 0.061 | 7   | 56.3   | <0.001** | 7   | 3.57  | 0.007   |
| G- PLFAs                   | 1   | 7.58  | 0.051 | 7   | 4.82   | 0.001**   | 7   | 3.45  | 0.009**   |
| G+/G- PLFAs                | 1   | 10.85 | 0.017* | 7   | 8.87   | <0.001** | 7   | 2.99  | 0.012*   |
| MBC                        | 1   | 0.06  | 0.818 | 7   | 5.16   | <0.001** | 7   | 0.36  | 0.940   |
| MBN                        | 1   | 1.03  | 0.339 | 7   | 6.07   | <0.001** | 7   | 0.12  | 0.999   |
| MBC/MBN                    | 1   | 0.16  | 0.707 | 7   | 12.42  | <0.001** | 7   | 2.44  | 0.035*   |
| Dissolved organic carbon   | 1   | 0.06  | 0.813 | 7   | 101.87 | <0.001** | 7   | 74.45 | <0.001** |
| Dissolved organic nitrogen | 1   | 68.46 | <0.001** | 7   | 7.21   | <0.001** | 7   | 6.59  | <0.001** |
| NH₄-N                      | 1   | 445.28 | <0.001** | 7   | 26.62  | <0.001** | 7   | 56.80 | <0.001** |
| NO₃-N                      | 1   | 62.06 | <0.001** | 7   | 11.37  | <0.001** | 7   | 18.97 | <0.001** |
| Invertase                  | 1   | 2.11  | 0.185 | 7   | 7.98   | <0.001** | 7   | 1.12  | 0.230   |
| Urease                     | 1   | 22.28 | 0.002** | 7   | 240.05 | <0.001** | 7   | 5.63  | 0.002** |
| Nitrate reductase          | 1   | 0.30  | 0.601 | 7   | 35.90  | <0.001** | 7   | 1.02  | 0.428   |
| Nitrite reductase          | 1   | 1.63  | 0.238 | 7   | 26.30  | <0.001** | 7   | 0.75  | 0.355   |

N. naphthalene treatment; T, sampling time. G+, gram-positive bacteria; G-, gram-negative bacteria; MBC, microbial biomass carbon; and MBN, microbial biomass nitrogen. Asterisks indicate significant (*P<0.05, **P<0.01) differences between the control and the naphthalene treatments over the whole of the experiment.

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Fig 1. Soil microbial respiration rates in subalpine forest soil of microcosms treated with and without naphthalene. Asterisks indicate significant (*P<0.05) differences between the treatments with naphthalene and without naphthalene at the same sampling time.

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Naphthalene effects on the soil microbial biomass

Sampling time had a significant \( P < 0.001 \) influence on the soil microbial biomass in the microcosms (Table 1). Tests for time trends showed a similar dynamic for the soil MBC and MBN and the ratio of MBC to MBN in both the naphthalene and control microcosms (Fig 3). There was a nonsignificant effect of naphthalene on the contents of MBC (\( F = 0.06, P = 0.818 \)), MBN (\( F = 1.03, P = 0.339 \)) and the ratio of MBC to MBN (\( F = 0.16, P = 0.707 \)) during the whole incubation period, but the interaction of naphthalene application and sampling time exerted a significant \( F = 2.44, P = 0.035 \) influence on the ratio of MBC to MBN (Table 1).

Naphthalene effects on the soil dissolved organic matter and inorganic N

The DOC, DON and inorganic N (\( \text{NH}_4^+ - \text{N} \) and \( \text{NO}_3^- - \text{N} \)) contents in both the naphthalene and control microcosms changed significantly \( P < 0.001 \) over time during the experiment. However, the DOC content was not significantly \( P > 0.05 \) affected by the naphthalene application (Table 1, Fig 4). Compared to the control, the naphthalene application triggered a sharp increase in DOC content in the first 3 days following the incubation and then decreased the content to a lower value that was maintained for most of the study (Fig 4A). Moreover, the contents of DON and inorganic N were not significantly \( P > 0.05 \) influenced
Fig 3. Soil microbial biomass contents in subalpine forest soil microcosms treated with and without naphthalene. The values represent the means ± the standard error (SE) (n = 5). Asterisks indicate significant (*P<0.05) differences between the treatments with naphthalene and without naphthalene at the same sampling time.

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Fig 4. Soil dissolved organic matter and extractable N contents in subalpine forest soil microcosms treated with and without naphthalene. The values represent the means ± the standard error (SE) (n = 5). Asterisks indicate significant (*P<0.05) differences between the treatments with naphthalene and without naphthalene at the same sampling time.

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by the naphthalene application in the first 17 days of the incubation but started to rapidly decline following the third naphthalene application on day 38 (Fig 4B–4D). In addition, the final contents of soil inorganic N were significantly different between the treatments and controls (Table 2). The absolute contents of soil inorganic N increased by approximately 30% in the controls, while the absolute contents of soil organic N in the treatments decreased by approximately 68% (Table 2).

**Naphthalene effects on the soil enzyme activity**

The soil enzyme activity was significantly affected \((P < 0.001)\) by the sampling time (Table 1), and the dynamics of these changes were similar to those found in the microcosms (Fig 5). The activities of invertase, nitrate reductase and nitrite reductase in both the treatments and controls were not significantly different (Table 1). Compared to the control, the naphthalene application decreased urease activity to a significantly lower level \((F = 22.48, P = 0.002)\) during

![Graphs showing soil enzyme activities](https://doi.org/10.1371/journal.pone.0217178.g005)

Fig 5. Soil enzyme activities in subalpine forest soil microcosms treated with and without naphthalene. The values represent the means ± the standard error (SE) \((n = 5)\). Asterisks indicate significant \((P < 0.05)\) differences between the treatments with naphthalene and without naphthalene at the same sampling time.

**Table 2.** Soil net ammonification \((\text{NH}_4^+\text{-N})\), nitrification \((\text{NO}_3^-\text{-N})\) and inorganic N \((\text{NH}_4^+\text{-N} + \text{NO}_3^-\text{-N})\) mineralization over the course of the experiment in subalpine forest soil microcosms treated with naphthalene and without naphthalene.

| Variables         | Without naphthalene |                    | Naphthalene |                    |
|-------------------|---------------------|--------------------|-------------|--------------------|
|                   | Initial (mg kg\(^{-1}\)) | End (mg kg\(^{-1}\)) | NMin (mg kg\(^{-1}\)) | Initial (mg kg\(^{-1}\)) | End (mg kg\(^{-1}\)) | NMin (mg kg\(^{-1}\)) |
| \text{NH}_4^+\text{-N} | 69.32±9.86          | 81.32±10.47        | 12.00±12.28  | 69.32±9.86          | 14.30±0.76b          | -55.03±7.67b          |
| \text{NO}_3^-\text{-N} | 32.65±2.92          | 50.91±8.34a        | 18.25±7.40a  | 32.65±2.92          | 19.50±3.45b          | -13.15±3.22b          |
| Inorganic N       | 101.98±9.04         | 132.23±13.32a      | 30.25±12.78a | 101.98±9.04         | 33.80±2.96a          | -68.18±8.75b          |

NM, net mineralization rate. The nitrogen content is based on the analysis of five individual microcosms from each treatment before and after 52 days. The values represent the means ± the standard error SE \((n = 5)\). Letters indicate significant differences \((P<0.05)\) based on Student’s t-test.

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most of the experiment (Fig 5B), and the interaction effect of naphthalene application and sampling time exerted a significant ($F = 5.63, P = 0.002$) influence on urease activity (Table 1).

**Discussion**

In recent decades, naphthalene has been widely used as a biocide to reduce or eliminate target groups of soil fauna in field experiments in the study of ecological functions of soil fauna [3, 10, 12, 17]. However, whether naphthalene has potential nontarget effects on soil microorganisms and nutrients has long been debated [3, 14, 15]. Previous soil and soil-litter microcosm studies have suggested that naphthalene might stimulate microbial respiration [14, 37], microbial biomass [15, 38], abundance and FAD fungal activity [17, 37]. Similarly, total microbial respiration in our microcosms increased following the application of naphthalene on days 17 and 38 (Fig 1). The reduced response in the first 10 days of application may be due to the restructuring of microbial populations under short-term environmental stresses [39, 40]. This finding corresponds to an obvious fluctuation in the ratios of fungi to bacteria, $G^+$ to $G^-$ bacteria and MBC to MBN in the naphthalene microcosm (Fig 2E and 2F, Fig 3C), which is in agreement with the results showing a reduction in the radial growth of fungal cultures after 6 days of naphthalene treatment [16]. The subsequent increase in microbial respiration rates was perhaps due to the consumption of naphthalene as a carbon source by the microbial community [3, 14, 38] and the increase in fungal abundance in the naphthalene microcosm (Fig 2B, Fig 6). Moreover, the promotion effect on microbial respiration rates differed with the results Blair et al. [14] and Margesin et al. [37] owing to varying soil types (Table 3). This result
suggests that naphthalene application in subalpine soil also represents an exogenous C source for soil microorganisms.

In field experiments, naphthalene addition showed negligible direct effects on the abundance of PLFAs and carbon dynamics in field soil [3] and the abundance of total or FDA active fungi in litterbags [14]. However, the opposite results indicated that naphthalene or other bioicide treatments significantly reduced the number of bacteria and FDA active fungi and the radial growth of fungi in microcosm experiments [14, 16]. Compared to the microcosm results [3, 14, 16], naphthalene application did not influence the overall microbial activities (Table 1), measured as soil microbial PLFA abundance (Fig 2) and microbial biomass (Fig 3), in our microcosm. This finding is consistent with a field experimental result [3]. Furthermore, Cotrufo et al. [3] and others [41, 42] noted that naphthalene-C might be substantially utilized by G+ bacteria, Actinobacteria and, to greater extent, G- bacteria. Although the overall naphthalene treatment did not reach a statistically significant effect on the abundances of G+ and G-
bacteria (Table 1, Fig 6), the treatment obviously suppressed the $G^+$ and $G^-$ bacterial abundances before the third naphthalene application on day 38 (Fig 2C and 2D). Simultaneously, the abundances of fungi and bacteria were stimulated significantly by the naphthalene applications at the end of the experiment (Fig 2A and 2B). These disparate results can be mainly explained as follows: (1) different methods of measuring microbial abundance might have caused the different responses of microbial abundance to the naphthalene application; (2) the removal of predation by soil animals had a greater positive effect than the negative effect caused by naphthalene on the fungi [17], and thus, an overall increase was observed (Fig 6); and (3) naphthalene offered a sufficient C source to maintain fungal growth in the naphthalene microcosm [37, 38]. In addition, the response of soil MBC to the naphthalene application (Fig 3) differed from the response observed by Xiong et al. [15]; the authors found that soil MBC was obviously increased by the same naphthalene application rate (100 g m$^{-2}$), which might be due to the higher soil organic matter content in the subalpine forest soil [19, 38] and differences in incubation times and temperatures [15].

The application of naphthalene and others biocides (e.g., triflumuron and profenofos) may introduce exogenous C and nutrients (e.g., N and P) as energy sources to microbes [15, 43] and subsequently stimulate or inhibit soil microbial activity and soil mineralization processes [44]. The naphthalene application significantly altered the soil nitrogen dynamics in our microcosms. The treatment suppressed increases in soil DON and inorganic N ($NH_4^+$-N and $NO_3^-$-N) and led to inorganic N net mineralization, which contrasted with the net immobilization result in the controls (Table 2, Fig 4B–4D). Similar results have been described by a previous study (Blair et al., 1989). The differences in DON dynamics might be due to the increase in fungal abundance following the third naphthalene application on day 38, which resulted in more utilization of DON by soil microorganisms (Geisseler et al., 2010). Additionally, the decreases in extractable $NH_4^+$-N and $NO_3^-$-N during the later stage of the incubation (Fig 4C and 4D) might be attributable to the greater abundances of fungi in the treatments, which promoted microbial immobilization of inorganic N or the translocation of inorganic N from soil to microorganisms [14, 45, 46]. Moreover, soil DOC derived from dead microorganisms under naphthalene stresses might have contributed to the pulsed increase in DOC during the first 3 days of incubation in the naphthalene microcosms (Fig 4A). These results partial confirm our hypotheses that naphthalene application might have a stronger nontarget effect on the soil biochemical properties of nitrogen cycling processes than on those of carbon cycling processes in alpine forest soil. However, it should be noted that ecosystem differences and a lack of plant absorption and utilization in microcosms can result in conflicting results for nontarget effects on nitrogen cycling processes (Table 3).

Soil enzyme activity is regarded as a key indicator of microbial activity under environmental stress [23, 47, 48]. In theory, naphthalene additions may affect extracellular enzyme activity in two ways [3]. First, reducing or removing targeted soil organism groups may directly affect other groups by altering the species-specific trophic behaviors or the interactions among these groups in the soil debris food chain [6, 49]. Second, the application of naphthalene might cause a nontarget effect on soil enzyme activity by stimulating soil respiration and microbial immobilization and abundance as well as available nutrients [3, 14]. In this study, although microbial metabolism (microbial respiration rate) and the available nutrients (organic and inorganic N) were significantly changed by the naphthalene treatments, soil invertase, nitrate reductase and nitrite reductase activities were not affected overall in the naphthalene microcosms, and similar dynamics were seen in the control (Table 1, Fig 5). Moreover, urease is an important hydrolytic biological enzyme for the conversion of organic nitrogen to available inorganic N ($NH_4^+$-N and others), whereas nitrate reductase and nitrite reductase transform $NO_3^-$-N and $NO_2^-$-N into $NH_4^+$-N [45, 50]. Studies have indicated that the production of
nitrogen-degrading enzymes, such as urease, is generally controlled by the microbial assimilation of \( \text{NH}_4^+ \) and \( \text{NO}_3^- \) in soil [45, 46]. Therefore, the greater fungal abundance in the naphthalene treatments (Fig 2B, Fig 6) might have stimulated the microbial assimilation of \( \text{NH}_4^+ \) and \( \text{NO}_3^- \) and then significantly repress urease activity (Table 1, Fig 5). This result was accompanied by decreases in inorganic N content after the third naphthalene application on day 38 in the naphthalene microcosms (Fig 4C and 4D). In addition, although urease activity was repressed by the microbial assimilation of inorganic N, an increasing trend was observed following the incubation (Fig 5B). A likely explanation is that the microbial assimilation of inorganic N facilitated the mineralization of soil DON (Fig 4B) by urease production in the soil in the naphthalene treatment (McCarty et al., 1992). These results further demonstrate that naphthalene application might exert substantial nontarget effects on the soil biochemical properties of nitrogen cycling processes in subalpine forest soil.

**Conclusions and implications**

In summary, this microcosm experiment explored the nontarget effects of naphthalene on soil microbial activities and soil nutrients by adding naphthalene to subalpine forest soil. Our results suggest that naphthalene application in subalpine soils also represents an exogenous C source for soil microbial respiration (Fig 6). The statistical analyses showed that naphthalene application did not affect the microbial activities overall, measured as soil microbial PLFA abundances and biomasses, or most enzyme activities during the whole incubation period. Overall, naphthalene application appeared to increase fungal abundance but had the opposite effect on bacterial abundance in the microcosms (Fig 6). However, the biocide application suppressed increases in DON, \( \text{NH}_4^+ \) and \( \text{NO}_3^- \) contents and urease activity and led to inorganic N (\( \text{NH}_4^+ + \text{NO}_3^- \)) net mineralization (Fig 6), which was contrary to the net immobilization result of the controls. Therefore, nontarget effects on soil nitrogen mineralization processes might occur when treating soil animals with naphthalene in a field experiment in subalpine forests. Caution should be taken in ascribing any changes in soil processes when using naphthalene to repel soil animals in field experiments. In addition, it should be acknowledged that there is a lack of uptake and turnover of aboveground vegetation in microcosms and whether this non-target effect exists in situ warrants further study. To improve the prediction of the potential nontarget effects of naphthalene application on soil biochemical properties in various ecosystems, ecosystem types, soil properties, soil-plant transformations, soil organism diversity and other factors associated with the soil biochemical cycle must be considered (Fig 6, Table 3).

**Supporting information**

S1 Dataset. Dataset file for the manuscript. The values represent the means and standard error (SE) of soil microcosms treated with and without naphthalene, respectively. (XLSX)

**Author Contributions**

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References
1. Swift MJ, Andren O, Brussaard L, Briones M, Couteaux MM, Ekschmitt K, et al. Global change, soil biodiversity, and nitrogen cycling in terrestrial ecosystems: three case studies. Glob Change Biol. 1998; 4: 729–743.
2. Paul EA. Soil microbiology, ecology and biochemistry. Academic Press; 2014.
3. Cotrufo MF, Soong J, Vandegeheuvel ML, Nguyen T, Denef K, Shaw EA, et al. Naphthalene addition to soil surfaces: a feasible method to reduce soil micro-arthropods with negligible direct effects on soil C dynamics. Appl Soil Ecol. 2014; 74: 21–29.
4. Huhta V. The role of soil fauna in ecosystems: A historical review. Pedobiologia. 2007; 50: 489–495.
5. Coleman DC, Whitman WB. Linking species richness, biodiversity and ecosystem function in soil systems. Pedobiologia. 2005; 49: 479–497.
6. Cragg RG, Bardgett RD. How changes in soil faunal diversity and composition within a trophic group influence decomposition processes. Soil Biol Biochem. 2001; 33: 2073–2081.
7. Moore JC, Berlow EL, Coleman DC, de Ruiter PC, Dong Q, Hastings A, et al. Detritus, trophic dynamics and biodiversity. Ecol Lett. 2004; 7: 584–600.
8. Brussaard LB, Pulleman MM, Ouedraogo E, Mando A, Six J. Soil fauna and soil function in the fabric of the soil food web. Pedobiologia. 2007; 50: 447–462.
9. Liu YW, Yang F, Yang WQ, Wu FZ, Xu ZF, Liu Y, et al. Effects of naphthalene on soil fauna abundance and enzyme activity in the subalpine forest of western Sichuan, China. Sci Rep. 2019; 9: 2849. https://doi.org/10.1038/s41598-019-39603-6 PMID: 30809005
10. González G, Seastedt TR. Soil fauna and plant litter decomposition in tropical and subalpine forests. Ecology. 2001; 82: 955–964.
11. Díaz S, Symstad AJ, Chapin FS III, Wardle DA, Huenneke LF. Functional diversity revealed by removal experiments. Trends Ecol Evol. 2003; 18: 140–146.
12. Wang SJ, Ruan HH, Wang B. Effects of soil microarthropods on plant litter decomposition across an elevation gradient in the Wuyi Mountains. Soil Biol Biochem. 2009; 41: 891–897.
13. Seastedt TR, Crossley Jr DA. Nutrients in forest litter treated with naphthalene and simulated throughfall: a field microcosm study. Soil Biol Biochem. 1983; 15: 159–165.
14. Blair JM, Crossley DA Jr, Rider S. Effects of naphthalene on microbial activity and nitrogen pools in soil-litter microcosms. Soil Biol Biochem. 1989; 21: 507–510.
15. Xiong YM, Shao YH, Xia HP, Li ZA, Fu SL. Selection of selective biocides on soil microarthropods. Soil Biol Biochem. 2008; 40: 2706–2709.
16. Newell K, Frankland JC, Whittaker JB. Effects on microflora of using naphthalene or X-rays to reduce arthropod populations in the field. Biol Fertil Soils. 1987; 3: 11–13.
17. Coleman DC, Crossley DA Jr, Ingham ER. Use of sulfamethoxazole-penicillin, oxytetracycline, carbafuran, carbaryl, naphthalene and temik to remove key organism groups in soil in a corn agroecosystem. J Sustain Agr. 1994; 4: 7–30.
18. Iuss Working Group Wrb. World reference base for soil resources 2014: International soil classification system for naming soils and creating legends for soil maps. World Soil Resources Reports FAO, Rome; 2014.
19. Tan B, Wu FZ, Yang WQ, Yu S, Liu L, Wang A, et al. Activities of soil oxidoreductase and their response to seasonal freeze-thaw in the subalpine / alpine forests of western Sichuan. Acta Ecol Sin. 2012; 32: 6670–6678 (in Chinese with English abstracts).
20. Chen CR, Condron LM, Davis MR, Sherlock RR. Effects of afforestation on phosphorus dynamics and biological properties in a New Zealand grassland soil. Plant Soil. 2000; 220: 151–163.
21. Lu RK. Soil and agro-chemical analytical methods. China Agricultural Science and Technology Press; 1999 (in Chinese).
22. Jones DL, Willett VB. Experimental evaluation of methods to quantify dissolved organic nitrogen (DON) and dissolved organic carbon (DOC) in soil. Soil Biol Biochem. 2006; 38: 991–999.
23. Tan B, Wu FZ, Yang WQ, He XH. Snow removal alters soil microbial biomass and enzyme activity in a Tibetan alpine forest. Appl Soil Ecol. 2014; 76: 34–41.

24. Brookes PC, Landman A, Pruden G, Jinkinson DS. Chloroform fumigation and the release of soil nitrogen: a rapid direct extraction method to measure microbial biomass nitrogen in soil. Soil Biol Biochem. 1985; 17: 837–842.

25. Vance ED, Brookes PC, Jenkinson DS. An extraction method for measuring soil microbial biomass C. Soil Biol Biochem. 1987; 19: 703–707.

26. Wang QK, Wang SL, Liu Y. Responses to N and P fertilization in a young *Eucalyptus dunnii* plantation: Microbial properties, enzyme activities and dissolved organic matter. Appl Soil Ecol. 2008; 40: 484–490.

27. Xiong L, Xu ZF, Wu FZ, Yang WQ, Yin R, Li ZJ, et al. Effects of snow pack on soil nitrogen transformation enzyme activities in a subalpine *Abies faxoniana* forest of western Sichuan, China. Chin. J Appl Ecol. 2014; 25: 1293–1299 (in Chinese with English abstracts).

28. White DC, Stair JO, Ringelberg DB. Quantitative comparisons of in situ microbial biodiversity by signature biomarker analysis. J Ind Microbiol. 1996; 17: 185–196.

29. He HY, Yang KJ, Li ZJ, Schädler M, Yang WQ, Wu FZ, et al. Effects of forest conversion on soil microbial communities depend on soil layer on the eastern Tibetan Plateau of China. PloS one. 2017; 12: e0186053. https://doi.org/10.1371/journal.pone.0186053 PMID: 28982191

30. Denef K, Roobroeck D, Wadu MCM, Lootens P, Booneck P. Microbial community composition and rhizodeposit-carbon assimilation in differently managed temperate grassland soils. Soil Biol Biochem. 2009; 41: 144–153.

31. Liu L, Gundersen P, Zhang T, Mo JM. Effects of phosphorus addition on soil microbial biomass and community composition in three forest types in tropical China. Soil Biol Biochem. 2012; 44: 31–38.

32. Ushio M, Balser TC, Kitayama K. Effects of condensed tannins in conifer leaves on the composition and activity of the soil microbial community in a tropical montane forest. Plant Soil. 2013; 365: 157–170.

33. Frosteåård Å, Tunlid A, Bååth E. Phospholipid fatty acid composition, biomass, and activity of microbial communities from two soil types experimentally exposed to different heavy metals. Appl Environ Microbiol. 1993; 59: 3605–3617. PMID: 16349080

34. Zelles L. Fatty acid patterns of phospholipids and lipopolysaccharides in the characterisation of microbial communities in soil: a review. Biol Fertil Soils. 1999; 29: 111–129.

35. Zogg GP, Zaik DR, Ringelberg DB, White DC, MacDonald NW, Pregitzer KS. Compositional and functional shifts in microbial communities due to soil warming. Soil Sci Soc Am J. 1997; 61: 475–481.

36. Tomberg K, Bååth E, Olsson S. Fungal growth and effects of different wood decomposing fungi on the indigenous bacterial community of polluted and unpolluted soils. Biol Fertil Soils. 2003; 37: 190–197.

37. Margesin R, Walder G, Schinner F. The impact of hydrocarbon remediation (diesel oil and polycyclic aromatic hydrocarbons) on enzyme activities and microbial properties of soil. Acta Biotechnol. 2000; 20: 313–333.

38. Margesin R, Labbe D, Schinner F, Greer CW, Whyte LG. Characterization of hydrocarbon-degrading microbial populations in contaminated and pristine alpine soils. Appl Environ Microbiol. 2003; 69: 3085–3092. https://doi.org/10.1128/AEM.69.6.3085-3092.2003 PMID: 12788702

39. Liu Q, Tanga, J, Liu X, Song B, Zhen M, Ashbolt NJ. Response of microbial community and catabolic genes to simulated petroleum hydrocarbon spills in soils/sediments from different geographic locations. J Appl Microbiol. 2017; 123: 875–885. https://doi.org/10.1111/jam.13548 PMID: 28763134

40. Sun GD, Du Y, Yin JX, Jiang YZ, Zhang DY, Jiang B, et al. Response of microbial communities to different organochlorine pesticides (OCPs) contamination levels in contaminated soils. Chemosphere. 2019; 215: 461–469. https://doi.org/10.1016/j.chemosphere.2018.09.160 PMID: 30336323

41. Denome SA, Stanley DC, Olson ES, Young KD. Metabolism of dibenzothiophene and naphthalene in *Pseudomonas* strains–complete DNA-sequence of an upper naphthalene catabolic pathway. J Bacteriol. 1993; 175: 6890–6901. https://doi.org/10.1128/jb.175.21.6890-6901.1993 PMID: 8226631

42. Kleemann R, Meckenstock R. Anaerobic naphthalene degradation by Gram-positive, iron-reducing bacteria. FEMS Microbiol Ecol. 2011; 78: 488–496. https://doi.org/10.1111/j.1574-6941.2011.01193.x PMID: 22066721

43. Sánchez ME, Estrada IB, Martínez O, Martín-Villacorta J, Aller A, Morán A. Influence of the application of sewage sludge on the degradation of pesticides in the soil. Chemosphere. 2004; 57: 673–679. https://doi.org/10.1016/j.chemosphere.2004.07.023 PMID: 15488930

44. Gonod LV, El Arfaoui A, Benoit P. Impact of spatial distribution of exogenous organic matter on C mineralization and isoproturon fate in soil. Soil Biol Biochem. 2016; 95: 180–188.
45. McCarty G, Shogren DR, Bremner JM. Regulation of urease production in soil by microbial assimilation of nitrogen. Biol Fertil Soils. 1992; 12: 261–264.

46. Geisseler D, Horwath WR, Joergensen RG, Ludwig B. Pathways of nitrogen utilization by soil microorganisms—a review. Soil Biol Biochem. 2010; 42: 2058–2067.

47. Sardans J, Penuelas J, Estiarte M. Changes in soil enzymes related to C and N cycle and in soil C and N content under prolonged warming and drought in a Mediterranean shrubland. Appl Soil Ecol. 2008; 39: 223–235.

48. Li XJ, Wang X, Zhang YY, Zhao Q, Yu BB, Li YT, et al. Salinity and conductivity amendment of soil enhanced the bioelectrochemical degradation of petroleum hydrocarbons. Sci Rep. 2016; 6: 32861. https://doi.org/10.1038/srep32861 PMID: 27597387

49. Trap J, Bonkowski M, Plassard C, Villenvae C, Blanchart E. Ecological importance of soil bacterivores for ecosystem functions. Plant Soil. 2016; 398: 1–24.

50. Roscoe R, Vasconcellos CA, Furtini-Neto AE, Guedes GA, Fernandes LA. Urease activity and its relation to soil organic matter, microbial biomass nitrogen and urea-nitrogen assimilation by maize in a Brazilian Oxisol under no-tillage and tillage systems. Biol Fertil Soils. 2000; 32: 52–59.