Impact of Three Bleaching Events on the Reef Resiliency of Kāneʻohe Bay, Hawaiʻi

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Coral bleaching events have been increasing in frequency and severity worldwide. The most prolonged global bleaching event began in 2014 and continued into 2017 impacting more reefs than any previous occurrence. Here we present the results of coral bleaching and mortality surveys conducted in Kāneʻohe Bay Oʻahu, Hawaiʻi and compare them to the only other widespread bleaching events to impact the main Hawaiian Islands in 1996 and 2014. Results from these surveys along with associated environmental factors were used to compare these events to gain a baseline understanding of the physical processes that influence localized bleaching dynamics under these extreme environmental conditions. Survey results show extensive variation in bleaching (1996–62%, 2014–45%, 2015–30%) and cumulative mortality (1996—<1%, 2014–13%, 2015–22%) between years. Bleaching prevalence was observed to decrease in certain reef areas across events, suggesting some acclimation and/or resilience, but possible increase susceptibility to mortality. Long-term monitoring sites show a similar temporal pattern of coral mortality and decline in coral cover, but revealed some reefs remained relatively un-impacted by consecutive high temperature events. Across the three bleaching events, we found that although circulation patterns can facilitate heating, the duration and magnitude of the high temperature event were the primary forcing functions for coral bleaching and mortality. Other localized primary drivers influencing water temperature such as irradiance, turbidity, and precipitation contributed to spatial variations. Recovery and resilience of this coral reef ecosystem is dependent on many factors including duration and magnitude of heating, resulting mortality levels, localized environmental factors in the bay, and coral species affected and their bleaching tolerances.

Keywords: coral reefs, resiliency, ecology and conservation, environmental monitoring, climate change impacts, coral bleaching

INTRODUCTION

Coral reefs throughout the world are undergoing significant ecological decline due to climate change (Aronson et al., 2002; Gardner et al., 2005; Norström et al., 2009; De’ath et al., 2012). Increases in anthropogenic atmospheric gases mainly from fossil fuel burning have resulted in significant increases in global sea surface temperatures (SST) (Sabine et al., 2004). Increased SST have resulted in more severe and more frequent coral bleaching events worldwide. 2014 marked the beginning of the longest global bleaching event on record, which continued for nearly three
and has affected more reefs than any previous worldwide bleaching event (Eakin et al., 2016). As warming continued, not only was 2015 the warmest year on record, but it also saw the single largest ocean temperature increase (0.16°C) within any given year (Heron et al., 2016).

The offshore subtropical waters of the Hawaiian Islands have experienced heating over the past 58 years (+1.15°C), which appears to have accelerated the frequency and severity of bleaching events throughout the archipelago (Hoeke et al., 2009; Bahr et al., 2015a). Previously recorded bleaching events (i.e., 1996, 2002, 2004) in the Hawaiian Archipelago were relatively short in duration and resulted in high recovery; however, reefs experienced unsurpassed bleaching on a statewide scale during the multi-year bleaching events in 2014 and 2015 (Bahr et al., 2015a).

The largest sheltered body of water in the main Hawaiian Islands, Kāne’ohe Bay, contains among the highest coral cover in the State with numerous patch reefs, a fringing reef, and a barrier reef. This intensively studied estuarine coral reef ecosystem has been impacted by humans for over 700 years (Bahr et al., 2015b). Drastic modifications in the terrestrial environment (e.g., urbanization, stream channelization, direct release of primary sewage into the bay for nearly 20 years (circa 1960–1979) decreased species diversity, increased eutrophication, and altered ecosystem structure away from a coral dominated ecosystem (Bahr et al., 2015b). By 1973, Kāne’ohe Bay was declared an ecosystem under stress until sewage diversion led to natural recovery of the coral reefs. The corals in Kāne’ohe Bay have shown resilience by their capacity to respond to the numerous perturbations by recovering rapidly. Whether this is a result of a reduction in genetic diversity or that acclimatization and/or adaptations have occurred, the population has shown high resilience. Climate change now poses a significant new threat to the integrity of this unique coral reef ecosystem. Kāne’ohe Bay has experienced initial indicators of climate change with more frequent localized storm events and flooding (i.e., freshwater kill events in 1965, 1988, 2014) and three major bleaching events (i.e., 1996, 2014, 2015). Thermally stressed corals are more susceptible to disease (Miller et al., 2009; Mydlarz et al., 2009; Burge et al., 2014), predation, bioerosion, and have reduced capacity for competition, growth, reproduction, and survival (Jokiel, 2004; Baker et al., 2008; Jones, 2008). Continued exposure to bleaching stress will eventually lead to reduced reef complexity, coral cover and biodiversity; therefore, providing lower quality habitat for fishes and marine invertebrates and consequently fewer ecosystem goods and services for dependent human communities (Munday et al., 2008).

Along with the rather unnatural history of Kāne’ohe Bay, these corals are currently living at temperatures (+1–2°C) and acidification (average pCO$_2$ levels ~500 µatm) regimes that will not be experienced for decades on open coastal reefs across the Hawaiian Archipelago (Bahr et al., 2015b). To determine whether these conditions increase the susceptibility of these corals to future anthropogenic and climate change stressors, we investigated the response and recovery of this coral reef ecosystem from recent bleaching events. Corals in the bay were closely monitored and documented during the previous 1996, 2014, and 2015 bleaching events. Previous work has documented the 1996 (Jokiel and Brown, 2004) and 2014 (Bahr et al., 2015a) bleaching events. The methodology and observer remained constant across all bleaching events. This presented a unique opportunity to examine all three bleaching events to determine differences in coral bleaching and mortality and their relationships with environmental factors. Furthermore, detailed long-term benthic monitoring surveys have been conducted by the Coral Reef Assessment and Monitoring Program (CRAMP) since 1999 to identify the controlling factors (i.e., natural and anthropogenic) contributing to bleaching variability across years, and/or recovery of the reefs in Kāne’ohe Bay (Brown et al., 2004). Together, these data reveal the impact of these bleaching events on the resiliency of this unique coral reef ecosystem. Here we present the survey findings from the 2015 bleaching event in Hawai’i and interpret those findings in the context of the two previously documented bleaching events. The objectives of this research were to: (1) describe the severity and extent of the 2015 bleaching event, (2) investigate the influence of the environmental drivers on bleaching variability across years, and (3) examine coral response and recovery during consecutive bleaching events to aid in understanding resiliency of the Kāne’ohe Bay reef ecosystem.

**MATERIALS AND METHODS**

**Bay-Wide Coral Bleaching and Mortality Surveys**

Documentation of the 1996 and 2014 bleaching events have been previously described in Jokiel and Brown (2004) and Bahr et al. (2015a), respectively. Here we described our detailed methods to quantify the extent of bleaching and associated mortality during the 2015 bleaching event. All bleaching surveys (1996, 2014, and 2015) were conducted using the same methodology described below.

Bay-wide coral surveys were conducted following the peak of the high temperature event for all three documented bleaching events in Kāne’ohe Bay, O’ahu Hawai’i (21°28′N; 157°48′W).

The goal of our bay-wide surveys was to detect change in coral cover following the high temperature event; therefore, site selection criteria for each survey year included the same depth (1–2 m), presence of coral, and spatial coverage. Surveys were conducted between 1 and 2 m on the reef flat and slopes were coral species identification and pigmentation level could be accurately estimated without discoloration or distortion with depth. Stations were surveyed on Leeward and Windward sides of patch reefs when possible to account for potential differences in coral cover. Extent of bleaching was defined by surveying fringing, patch, and barrier reefs within Kāne’ohe Bay across bleaching years (i.e., 1996, 2014, and 2015). To define bleaching prevalence, rapid assessments of total coral cover, species composition (% of live coral cover), and coral condition (i.e., normal, pale, bleached, dead expressed as % of live coral cover) at each site were conducted using a visual estimate technique at the same sites using the same observer (Bahr et al., 2015a). Individual coral colonies were not quantified, instead conditions...
of live coral area (100 m$^2$) were estimated. These assessments were conducted from a boat at the surface under clear, calm conditions using a clear viewing box and/or snorkeling along the reef edge and across reef flats during a 5-min observation period that covered an area of 100 m$^2$ (Bahr et al., 2015a). This method was previously compared to eight other quantitative methods (i.e., quadrat, random, point intercept transect, CRAMP assessments, video transect, towed-diver, photographic transect, NOAA ground truth) and produces comparable results in total coral cover estimates (Jokiel et al., 2015). This visual estimate technique allowed us to rapidly evaluate bleaching severity and extent across the large sampling area but may show higher variance. Coral condition was defined by visual assessment of pigmentation: “normal” (fully pigmented), “pale” (obvious pigmentation loss but retention of color), “bleached” (loss of all color), and recently “dead” skeleton as a result of bleaching mortality. These coral condition classifications follow the same methodology and classification originally established by Jokiel and Coles (1974) and were used in previous bleaching surveys (i.e., 1996 and 2014 bleaching events) at the same reef locations by the same observers (Bahr et al., 2015a).

**The 2015 Bleaching Event**
Bay-wide surveys ($n = 24$) were conducted ~25 days following peak temperature on 16 October 2015 in areas of high coral cover at a depth of 1–2 m. These sites were revisited after the 2014 surveys to confirm bleaching occurrence covering the spatial extent of the Bay. Sample intensity was lower in Oct 2015 and is accounted for in our statistical approach. Coral recovery surveys were conducted on 8 December 2015 to assess coral mortality and pigmentation on the reefs previously surveyed ($n = 103$ sites) using identical methods and observer. Additional surveys of coral condition, coral cover, and species composition were conducted on 11 November 2016 to determine the degree of recovery from the 2015 bleaching event ($n = 100$ sites). Baywide coral surveys from 1996, 2014, and 2015 were compared to investigate variation in bleaching response across documented bleaching events. Results from bay-wide surveys were then compared to long-term CRAMP surveys (shallow sites, 2m depth) to investigate the effects of these high temperature events on long-term coral cover changes.

**Long-Term Coral Reef Assessment and Monitoring (CRAMP) Sites**
The Hawai‘i CRAMP was developed in 1998 to describe changes on reefs over time in relation to natural and anthropogenic factors. This network of 62 permanent reef stations at 31 sites are stratified by depth (i.e., shallow ~3 m and deep ~10 m) across the main Hawaiian Islands. Three of these sites are located within Kāne‘ohe Bay, at two depths (2 and 8 m). The shallow sites are at comparable depths to the bay-wide surveys, thus the shallow sites were used to assess the effects of bleaching impacts on long-term coral cover trends. These sites are located in three different sections of the bay, the north bay (patch reef 44, Ka‘alaea; flow regime 4), central bay (patch reef 17, He‘eia; flow regime 5) and south bay (Moku o Lo‘e Island, flow regime 6) (Figure 1; Table 1). At each station, 10 randomly selected 10 m transects and five photo quadrats within a 200 m$^2$ area are permanently marked with short stainless steel pins. Non-overlapping digital images estimate benthic coverage using the software program PhotoGrid (Bird, 2001). Percent coral cover, coral species richness, and coral diversity are calculated. The methodology was designed with high statistical power to distinguish an absolute change of 10% in coral cover annually within a site with statistical power increasing over time. CRAMP methodology is described in further detail in Brown et al. (2004). Coral surveys at long-term monitoring sites in 2012 and 2016 serve to provide benchmarks for the reef status before and after the consecutive bleaching events. From these data, we can deduce changes in coral cover in relation to the high temperature event rather than potential natural oscillations in coral cover. Further, CRAMP provides supporting evidence of the presence and degree of mortality and the spatial pattern documented in the bay-wide surveys.

**Meteorological Data**
Meteorological data [i.e., photosynthetically active radiation (PAR), ultraviolet radiation, seawater temperature, precipitation, and wind strength and direction] were monitored continuously throughout the bleaching events (1996, 2014, 2015) at an automated weather station located on Moku o Lo‘e at the Hawai‘i Institute of Marine Biology (HIMB) (http://www.pacioos.hawaii.edu/weather/obs-mokuole/) (Figure 1). To investigate the influence of the environmental drivers on bleaching variability, meteorological data collected between Aug 1 and Nov 1 were compared across bleaching years (1996, 2014, and 2015).

Due to the unique bathymetry of the bay, water circulation and residency times vary between ~1 day in north bay to up to 1–2 months in south bay (Lowe et al., 2009). Continuous in situ water temperatures were recorded at 1–2 m depth at seven locations characterizing four flow regimes during the 2015 bleaching event (Table 1). Water temperatures were recorded in the following flow zones: Zone 2, on reef flat near Kapapa Island; Zone 4, on northern fringing reef near Channel Marker 12; Zone 5, on patch reefs 12, 19, 31; Zone 6, at three locations including fore reef slope of the HIMB reef adjacent to weather station, and on the reef flat of the south bay fringing reef near Kokokahi pier (Figure 1). Temperature at the HIMB weather station was recorded using a recording thermistor thermometer (Cole-Parmer Instrument Co., Vernon Hills, Illinois, USA). All other in situ water temperature measurements (i.e., Kapapa isle, Marker 12, PR 12, PR 19, PR 31, and Kokokahi) were recorded at 15-min intervals using replicate HOBO Water Temperature Pro v2 Data Loggers (Onset, MA, USA). All temperature loggers were placed in areas of well-developed coral communities in 1–2 m of water. The loggers were secured in 6′′ × 12′′ hand-poured concrete structures to mimic the benthic substrate, provide weight, and protect loggers from solar irradiance and associated heating (Bahr et al., 2016b).

Differences in cumulative heating between bleaching years (1996, 2014, and 2015) were calculated by adding number of degrees above optimal mid-day temperatures of 27°C (defined by Jokiel and Coles, 1990) using seawater temperatures from the HIMB weather station. Daily mid-day temperatures above 27°C
were summed to determine the duration and extent of cumulative heating up to the onset of bleaching during the three documented bleaching events.

**Statistical Analysis**

To characterize localized temperature regimes during the 2015 bleaching event, *in situ* seawater temperatures (mean mid-day...
temperatures 11:00–14:00) during the peak (above 27°C) of the 2015 bleaching event (1 Aug-25 Sept 2015) were analyzed using a one-way ANOVA. Post-hoc Tukey tests were used to determine which sites differed from the other.

Bay-wide assessment surveys from 1996, 2014, and 2015 were compiled to investigate the influence of environmental drivers on bleaching prevalence and severity in Kāne‘ohe Bay.

A Generalized Linear Mixed-effects Model (GLMM) was conducted to investigate bleaching prevalence (% live coral) across bleaching years using October bleaching surveys. Year (i.e., 1996, 2014, 2015), flow regime, and total coral cover (%) were included as predictor variables. The GLMM allows and accounts for analysis assuming unequal group variance and analysis of longitudinal data with unequal time points and missing time points. Collinearity between candidate predictor variables were assessed with Pearson correlation coefficients (<0.7) and variance inflation factors (<5). GPS location (i.e., latitude and longitude) of each surveyed area was included in the GLMM as a repeated covariance spatial isotropic spherical structure to account for spatial autocorrelation. A Principal Component Analysis was used to identify the environmental factor drove coral bleaching prevalence across bleaching years.

Average coral mortality (including zeros) by survey month during each bleaching event was compiled and analyzed using a three-way ANOVA with fixed factors of survey month, bleaching event year, flow regime, and the interaction between year and flow regime. Tukey post-hoc analyses were used to investigate differences across factors. Average daily mid-water temperatures (°C) (2 m depth), daily wind (mph), cumulative daily Photosynthetically Active Radiation (PAR) (µmol photons m⁻²), and log transformed total daily rainfall (mm) were analyzed using a GLMM by year with day as a random effect. All meteorological variables were recorded at the HIMB weather station on Moku o Lo‘e.

To understand the implications of the high temperature events, comparisons of changes in coral cover at long-term CRAMP sites were analyzed using a two-way ANOVA with fixed factors of site and year and their interaction. Preplanned contrasts were used to analyze changes in coral cover before bleaching (i.e., 2012) and after (i.e., 2016). All statistical analyses and descriptive statistics were conducted using JMP Pro 12 (SAS Institute Inc., USA).

**RESULTS**

**2015 Bleaching Event**

Mean mid-day seawater temperatures varied spatially within Kāne‘ohe Bay during the peak of the bleaching event (1 Aug-25 Sept 2015) [One-way ANOVA; \( F_{(6, 39)} = 32.3150; p < 0.0001 \)]. Temperatures were at least 0.60°C warmer in the sluggish waters in south bay (Kokokahi Pier; 29.78 ± 0.11°C) compared other sites. Lower water temperatures were recorded in central bay (patch reef 19, 28.66 ± 0.08°C; patch reef 12, 28.61 ± 0.07°C), Moku o Lo‘e (28.49 ± 0.08°C), and the barrier reef site, Kapapa (28.40 ± 0.09°C; Figure 2).

Bay-wide coral surveys in Oct 2015 determined 46% of surveyed corals had signs of bleaching and paling and 13% showed recent mortality due to bleaching. Detailed surveys by location revealed the highest levels of bleaching and paling in the north bay (70%) with 18% mortality, while the highest mortality was observed in the south bay (28%) accompanied by 60% bleaching and paling. The lowest mortality (4%) and bleaching/paling (38%) levels were observed in the central bay.

**Comparison of Documented Bleaching Events in Kāne‘ohe Bay, HI**

Results of the 1996 and 2014 bleaching events have been previously documented in Jokiel and Brown (2004) and Bahr et al. (2015a). These data were compiled with the 2015 surveys to investigate differences in bleaching prevalence and mortality across documented events.

**Bleaching Prevalence**

Bleaching prevalence varied by year (\( p = 0.0002 \)) and by flow regimes (\( p = 0.0003 \)) (Table 2). Bleaching prevalence was ~27% higher in 1996 (62.00% ± 15.42) in comparison to 2014 (45.28% ± 1.91; \( p = 0.0138 \)) even though bleaching was restricted to the inner bay and fringing reef in 1996, while bleaching extended throughout the bay to the barrier reef in 2014. Two consecutive years of bleaching revealed bleaching levels to be 15% lower in 2015 than in 2014 (Figure 4) (Supplementary Table 1). The highest bleaching prevalence was observed in zone 5 (48.93 ± 2.36%), zone 6 (45.24 ± 4.49%), and zone 4 (44.33 ± 4.82%).

**TABLE 1** | Descriptions of residency times and flow regimes in Kāne‘ohe Bay, Hawai’i by Lowe et al. (2009).

| Flow regime       | Dominant physical force     | Reef type                     | Residence time (days) |
|-------------------|-----------------------------|-------------------------------|-----------------------|
| 1                 | Wave                        | Barrier reef                  | <1                    |
| 2                 | Wave                        | Barrier reef                  | 4–10                  |
| 3                 | Tide                        | Barrier reef                  | 3–7                   |
| 4                 | Wave                        | Patch and fringing reef       | <1                    |
| 5                 | Wave                        | Patch and fringing reef       | 10–20                 |
| 6                 | Tide                        | Patch and fringing reef       | >30                   |

By December 2015, 2 months following the initial survey, high levels of recovery (~81% normal pigmentation) were observed (Figure 3e). Nearly 9% of surveyed corals in the bay exhibited signs of paling or bleaching, and recent mortality was estimated at 8%. A large proportion of coral mortality was attributed to the barrier reef (~56%). Recovery was delayed in the south bay where 25% of the surveyed corals remained bleached and/or pale in pigmentation. The highest recovery and lowest mortality rates (<2%) were observed in central bay. Normal pigmentation levels had returned by Nov 2016 (>99%) (Figure 3f).

Total live coral cover of surveyed area in Kāne‘ohe Bay (mean ± SE) was estimated at 59.17 ± 5.28% in Oct 2015. Higher total live coral cover was estimated during recovery surveys in December 2015 (68.95 ± 2.20%) and decreased again in November 2016 surveys (54.35 ± 1.98%; Supplementary Table 1).

### References

- Bahr et al. (2015a)
- Lowe et al. (2009)
Bleaching prevalence was lower in areas more influenced by oceanic waters (zone 1, 29.0 ± 11.87%; zone 2, 29.0 ± 9.83%; and zone 3, 17.5 ± 4.55%) across years (Figure 4B). Bleaching prevalence did not vary by total coral cover ($p = 0.1429$; Table 2).

Coral Mortality
Average coral mortality increased during the documented bleaching events ($p < 0.0001$) with high mean mortality in 2015 (16.07 ± 1.53%), less in 2014 (5.58 ± 1.42), and lowest in 1996 (<1%) (Supplementary Table 1; Table 3). Coral mortality varied by flow regime within years (year*flow, $p = 0.0005$). High mortality levels were observed in 2014 in flow regime 4 (28.81%) due to a freshwater flooding event that interacted synergistically with the high temperature event (Bahr et al., 2015a). Mortality during 2015 was similar across all flow regimes with the highest in flow regime 2 (39.16 ± 4.91%) (Table 3; Figure 5). Cumulative mortality (sum of both October and December surveys) was lowest in 1996 (<1%) and highest in 2015 (21.66 ± 5.32%).

Environmental Factors
Further investigation of environmental drivers (i.e., temperature, irradiance, wind, and precipitation) during the height of bleaching (1 Aug-1 Nov) revealed significant differences between bleaching years. Average daily mid-day temperatures recorded at 2 m at the HIMB weather station varied between bleaching events [GLMM; $F_{(2, 184)} = 7.62$; $R^2 = 0.31$; $p = 0.0007$]. 1996 had significantly higher temperatures compared to the 2015 event (Figure 6A). Cumulative daily solar (PAR) input was higher in 2014 than in 2015 [Mixed Model; $F_{(2, 184)} = 4.28$; $R^2 = 0.20$; $p = 0.0153$]. Similar levels of PAR were observed in 1996 (8,019.59 ± 122.69 µmol photons m$^{-2}$) compared to the 2014 (8,409.38 ± 214.92 µmol photons m$^{-2}$) and 2015 (7,616.33 ± 247.38 µmol photons m$^{-2}$) events (Figure 6B).
Long-term CRAMP sites were surveyed before (i.e., 2012) and following (i.e., 2016) consecutive bleaching events. In 2016, the highest coral cover was observed in central bay (He‘eia, 60.66 ± 3.61%) and north bay (Ka‘alaea, 49.13 ± 13.99%). The south bay site, Moku o Lo‘e, has significantly lower coral cover compared to the other shallow CRAMP sites in the bay (28.35 ± 2.53%; Figure 7A). Over the 4-year period (2012–2016), significant decline in coral cover occurred at shallow (2 m) sites in central bay, He‘eia (−19.7%; p = 0.0015), and south bay, Moku o Lo‘e (−42.2%; p < 0.0001; Figure 6). Coral cover did not significantly change over this 4 year period at the shallow north bay site (Ka‘alaea, p = 0.2640) [Two-way ANOVA; F(11, 119) =55.64; p < 0.0001] (Figure 7). Total coral cover was statistically similar among all CRAMP deep sites (8 m) in 2016 (Ka‘alaea = 6.45 ± 4.15%; He‘eia = 10.53 ± 2.06%; Moku o Lo‘e = 9.67 ± 2.34%) [One-way ANOVA; F(2, 29) = 1.2131; p = 0.3129] (Figure 7B) and did not change in response to the consecutive bleaching events [Two-way ANOVA; F(11, 119) = 55.64; p < 0.0001] (Figure 7).

**DISCUSSION**

The occurrence and severity of mass coral bleaching has increased dramatically over the last two decades with almost every reef region in the world suffering extensive bleaching and mortality (Hughes et al., 2017). The goals of this research were to (1) describe and compare the severity, extent, and spatial variability of bleaching during documented bleaching events and (2) examine recovery at long-term monitoring sites to aid in understanding the resiliency of Kāne‘ohe Bay, HI.

The first coral bleaching event in Kāne‘ohe Bay occurred during 1996 where high levels of bleaching but little mortality occurred. This event was followed by more severe events in 2014 and 2015. Patterns of bleaching and mortality in the 1996 event did not continue during the 2014 event (Figure 3). Likewise, the patterns observed in 2015 diverged from the previous two events, with higher levels of coral mortality. Our data suggests that the unexpected trends of bleaching and mortality within and across documented bleaching events were due to several factors including: (1) Localized patterns of water circulation and heating within bleaching events (Figure 2), (2) Variation in irradiance due to different levels of cloud cover and temporal variation in turbidity (Figure 6B), (3) Differences due to seasonal timing and duration of the warm water events (Figure 8), and (4) Influence of localized fresh water flood events (2014) interacting with temperature induced bleaching (Figure 3B).
During the 2015 bleaching, reefs across the Hawaiian Islands experienced different Degree Heating Weeks (DHW). Among the highest were on the west side of Hawai‘i Island, where surveyed corals experienced high levels of bleaching (up to 86%) and mortality (50%) after 18 DHW (Kramer et al., 2016). Conversely, after 12 DHW Kāne‘ohe Bay had 22% mortality, while the Hanauma Bay Nature Preserve (HBNP) on O‘ahu experienced high levels of bleaching (47%) and mortality (9.8%) after 10 DHW (Rodgers et al., 2017).

The 2015 Bleaching Event

Results of the 2015 bleaching event showed the influence of water circulation and localized heating on coral bleaching and mortality patterns inside Kāne‘ohe Bay. Historically, circulation patterns
in the bay have been described as the primary factor controlling distribution of coral species (Smith et al., 1981; Bahr et al., 2015b). Oceanic waters enter the bay over the barrier reef circulating cool water to adjacent patch reefs in central bay (Jokiel, 1991; Bahr et al., 2015b). The reef areas influenced by the movement of this cool oceanic water and low residency times had the lowest bleaching and mortality and highest recovery during the 2015 event. Net transport of water out of the bay occurs through the NW channel and the central Sampan channel (Figure 1). Although north bay is also characterized by high circulation, movement of warm water through the NW channel exacerbated bleaching (70%) and mortality (18%) on north bay reefs. These reefs have consistently had among the highest bleaching levels during the previous documented events (Figure 3). South bay has flow patterns anomalous to the rest of the bay with high residency times (~8 to > 30 days) and heavy riverine influence (Table 1). The high residency time and restricted circulation patterns facilitated localized heating (+0.65°C) in south bay and delayed recovery with high mortality rates (Figures 2, 3F). Although temperature was determined to be the driving factor of bleaching, reduced light penetration via wind-induced mixing and resuspension of sediments into the water column may explain lower than expected bleaching prevalence in south bay (Smith et al., 1973).

### Comparison of Documented Bleaching Events in Kāneʻohe Bay, Hawaiʻi

Different patterns of warming (i.e., timing and duration) greatly influenced severity and extent of bleaching in Kāneʻohe Bay across documented events (Figures 3, 8). Warm water approached the Hawaiian Islands from the north in 2014, while in 2015 warm water advanced from the south resulting in twice the duration of elevated temperatures as previous events with a maximum 6 DHW in 1996, 5 DHW in 2014, and 12 DHW in 2015 for Kāneʻohe Bay (Table 4). Seasonal timing also varied across documented events. Elevated seawater temperature began in late August in 1996 and in late September in 2014 (Figure 8). Once average water temperatures exceeded 27°C, initial signs of bleaching occurred after 28 days in 1996 (cumulative heating of 32°C), 62 days in 2014 (cumulative heating of 44°C), and just after 24 days in 2015 with the lowest cumulative heating (14°C) (Figure 9). Therefore, bleaching onset in 2015 occurred earlier (24 days) under lower cumulative heat stress (14°C). During the 1996 event, bleaching was restricted to the fringing and patch reefs. Conversely, bleaching during the 2014 and 2015 events extended throughout the bay onto the barrier reef. The severity of the high temperature event was lowest in 1996, where little to no mortality was recorded (Jokiel and Brown, 2004). In 2014, a freshwater kill event a month prior to the high temperature event, contributed significantly to observed mortality (nearly 60%). Areas not impacted by the freshwater flooding event experienced little to no mortality (<1%) (Bahr et al., 2015a; Figure 3b). The largest cumulative mortality (21%) due to high temperature was recorded in the most recent 2015 bleaching event in which warming began sooner, and corals bleached under less cumulative heating and lower irradiance regimes but experienced prolonged high SSTs at 12 DHW.

### Changes at Long-Term CRAMP Sites

Our long-term CRAMP sites provide a record of change in coral cover across the three bleaching events that provides supporting evidence of the validity of our mortality results within Kāneʻohe Bay (Rodgers et al., 2015). Significant increases in coral cover were reported at Moku o Loʻe shallow (south), Heʻeaia shallow (central), and Kaʻalaea deep (north) between 1999 and 2012. During the period of the more severe bleaching events in 2014 and 2015, significant declines in coral cover were reported at the shallow site in south bay (~42%) and the shallow site in central bay (~20%). Conversely, coral cover at the shallow north bay site (Kaʻalaea) and all three deeper sites (8 m) remained relatively stable. This is in concert with our bay-wide coral bleaching surveys conducted in 2014 and 2015. Additionally, bleaching prevalence was observed to decrease in certain flow regimes (4–6) across bleaching events, which may suggest some acclimation or...
Bahr et al. Bleaching Impacts on Reef Resiliency

**FIGURE 8** | Seasonal timing of 1996 (blue), 2014 (red), and 2015 (black) bleaching events. Daily mid-day (11:00–14:00) temperatures between 1 Aug and 1 Nov are shown for each year. Temperature data from 1993 is included as a non-bleaching reference year (yellow, dashed line). Arrows indicate when bleaching began. Temperature data collected from Hawai‘i Institute of Marine Biology Weather Station (http://www.pacioos.hawaii.edu/weather/obs-mokuoloe).

**TABLE 4** | Summary comparison of coral response and environmental parameters for all documented bleaching events in Kāne‘ohe, O‘ahu, Hawai‘i.

|                          | 1996         | 2014         | 2015         |
|--------------------------|--------------|--------------|--------------|
| Coral Response           |              |              |              |
| Bleaching (%)            | 62 ± 15.42   | 45.28 ± 1.91 | 29.58 ± 4.42 |
| Cumulative mortality (%) | 0.11 ± 0.11  | 12.76 ± 3.5  | 21.66 ± 5.32 |
| Environmental Parameters |              |              |              |
| Degree heating weeks (DHW)| 6           | 5            | 12           |
| Cumulative heating (°C)  | 32           | 44           | 14           |
| Midday temperature (°C)  | 28.16 ± 0.10 | 27.88 ± 0.07 | 27.65 ± 0.13 |
| Cumulative daily PAR (μmol m⁻²) | 8,019.59 ± 122.69 | 8,409.38 ± 214.92 | 7,616.33 ± 247.38 |
| Avg daily wind           | 10.13 ± 0.36 | 9.25 ± 0.35  | 9.56 ± 0.45  |
| Total daily rainfall (mm)| 0.06 ± 0.03  | 3.05 ± 0.89  | 4.34 ± 1.07  |

Mean ± SE environmental parameters between Aug 1-Nov 1 for each year. Cumulative mortality is the sum of average mortality in the October and December surveys.

reef resilience to bleaching in these areas, but possible increased susceptibility to mortality. The large mortality coral observed in 2014 and 2015 have contributed to the decline in coral cover at these long-term monitoring sites.

CRAMP surveys conducted in 2012 revealed overall coral cover and diversity in Hawai‘i have remained relatively stable since the initial survey in 1998 (Rodgers et al., 2015). During this period, coral cover on reefs in the Caribbean and other regions declined by as much as 50% due to climate change related bleaching events, increased storm damage and lowered coral growth (Wilkinson, 2004). Moreover, the recent longest, most widespread, and possibly the most damaging coral bleaching event on record had a significant impact on Australia’s Great Barrier Reef (Hughes et al., 2017). While some reef areas in Kāne‘ohe Bay remained relatively un-impacted by the consecutive bleaching events.

Offshore sea surface water temperatures in Hawai‘i have increased by approximately 1.15°C over the past six decades (Bahr et al., 2015a). Since the establishment of bleaching thresholds in 1974 (Jokiel and Coles, 1977), there has been a regional increase in temperature of over 0.8°C with three bleaching events documented in Kāne‘ohe Bay (1996, 2014, 2015). Reef recovery following disturbance has been documented at other CRAMP sites in the state (Rodgers et al., 2015); however, with projected increase in severity and frequency of bleaching events, we may see more significant declines in coral cover in the near future.

Recovery and resilience after these major disturbances not only depends on the prevailing environmental conditions but also the coral species affected. While Kāne‘ohe Bay holds some of the highest coral cover (>90%) in the Main Hawaiian Islands,
the majority of its reefs are comprised of two reef building species: *Porites compressa* and *Montipora capitata* (Bahr et al., 2015b). Previous experimental work has determined the endemic coral, *P. compressa*, to be the most susceptible to future climate change scenarios with the largest decreases in calcification and high partial mortality (Bahr et al., 2016c). A main factor in *M. capitata*'s resilience is its ability to increase heterotrophic feeding under high temperature conditions meeting all dietary requirements without their symbiotic zooxanthellae (Grottoli et al., 2006; Rodrigues and Grottoli, 2007; Hughes and Grottoli, 2013). Cryptic species (*Fungia scutaria* and *Leptastrea purpurea*) were observed to be more resilient to climate change. These species-specific responses correlate with the bleaching susceptibility hierarchy described during the bleaching events (Jokiel and Brown, 2004; Bahr et al., 2015a). Shifts in optimal growth rates have been documented and estimated in the second dominant species, *M. capitata*. By mid-century, growth rates are expected to decline by half with further reductions in growth at the end of the century with those corals exposed to high irradiance levels (e.g., reef flats) expected to have negative growth (Bahr et al., 2016a). With these reductions, shifts in seasonal growth patterns are expected. Currently, the highest growth rates occur in the late summer (Jokiel and Coles, 1977). As we progress to mid-century, maximum growth rates will shift to spring, and at the end of the century to winter due to changes in water temperature projections (Bahr et al., 2015a). The primary factor controlling coral populations in the future will be coral mortality caused by short-term high temperature excursions (i.e., bleaching events); however, localized environmental factors as well as differences in species composition may reveal resilient coral communities.

**CONCLUSION**

Results of this study reveal duration and magnitude of heating are the primary factors controlling the severity and extent of bleaching in Kāne'ohe Bay. Circulation patterns in the bay facilitate localized heating and influence spatial bleaching and its associated mortality. Sectors within the bay that were influenced by cool oceanic waters had low coral bleaching and mortality and high recovery rates. Conversely, restricted water movement facilitated localized heating and therefore increased bleaching and associated mortality. Other localized environmental factors that can influence water temperature (i.e., irradiance, turbidity, rainfall) may also reduce or amplify bleaching and explain spatial variations in bleaching and mortality. Long-term monitoring sites revealed significant declines in coral cover at some sites while others recovered from the consecutive bleaching events. Having an established network of long-term monitoring sites with associated temperature loggers allows for a reliable baseline to evaluate both natural and anthropogenic factors that contribute to the stability, decline and/or recovery of coral reefs and direct targeted management actions. Further, understanding of the driving factors that influence spatial variability in coral bleaching and mortality can assist managers in identifying areas to focus efforts and resources.

**AUTHOR CONTRIBUTIONS**

KB conceived and designed the experiments and surveys, performed the surveys, analyzed the data, wrote the paper, prepared figures and tables. KR performed and helped design the surveys, analyzed the data, reviewed drafts of the paper.
PJ conceived and designed the experiments, provide intellectual insights, and reviewed drafts of the paper.

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**SUPPLEMENTARY MATERIAL**

The Supplementary Material for this article can be found online at: https://www.frontiersin.org/articles/10.3389/fmars.2017.00398/full#supplementary-material

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**Conflict of Interest Statement:** The authors declare that the research was conducted in the absence of any commercial or financial relationships that could be construed as a potential conflict of interest.

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