WATER BEETLES: SPECIES DIVERSITY AND RESPONSE TO PHYSICO-CHEMICAL PARAMETERS IN DIFFERENT FRESHWATER HABITATS IN SAMTSE, BHUTAN

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Abstract

Bhutan is located between the two zoogeographic regions which are assumed to be rich in diversity of water beetles. The study aimed to determine the diversity and distribution pattern of water beetles in different freshwater habitats in Namgaychholing under Samtse district, Bhutan. The study also analysed the association of water beetles with the physico-chemical parameters. The data were collected from four different habitats with sample plot of 3 m × 1 m for a standard time of 1 hour in every plot. A total of 30 species belonging to 7 families were recorded of which Dytiscidae (n=626, RA=51) was the most abundant family and Lacconectus basalis (n=416, RA=34.13) was the most abundant species. The overall Shannon diversity index of the study area was $H' = 2.48$. Amongst the different habitats, marshy area had the highest Shannon diversity index ($H' = 2.45$) and stream had the lowest ($H' = 1.82$). A Kruskal Wallis test on diversity among different habitats showed no significant difference ($p<.05$). Through a cluster analysis, the species composition between the river and waterlogged showed 100% similarity index. Canonical correspondence analysis showed temperature as a main determining factor affecting the distribution of water beetles. With broad families encompassing the water beetles, similar study on water beetles in different regions of the country is recommended with particular emphasis on different families. Moreover, Bhutan falls in junction of Palearctic and Oriental Zoogeographic regions, which are known to be rich in diversity of water beetles. Hence, study in high-altitude lakes, which are assumed to be rich in diversity of water beetles are recommended despite social beliefs to keep lakes isolated and undisturbed.

Keywords: Marshy area; pH; Temperature; Water beetles

DOI: http://dx.doi.org/10.3126/ije.v10i1.38404

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Introduction

Coleoptera is the largest order under the class Insecta (Jach and Balke, 2008). The aquatic beetles are found to inhabit in almost all kinds of aquatic habitats such as rivers, springs, lakes, ditches, puddles, phytotelma, seepages and ground water. They are known to survive even when trapped in ice and high saline water (Gerdes et al., 1985). Beetles occur on all continents except Antarctica. Water beetles are found in all biogeographic regions and the highest diversity is found in the tropics. The smallest water beetles is less than 1 mm and the largest ones are more than 5 cm long. (Jack and Balke, 2008). Water beetles can either be herbivores, predators, or scavengers (Epler, 1996).

Water quality plays a key role in species richness of aquatic beetles; therefore, they could be considered as an indicator organism for ecosystem health of freshwater (Dong et al., 2014; Ghanem et al., 2018). The main physico-chemical factors that affect aquatic environments are temperature, discharge, dissolved oxygen (DO), pH, nutrients and electrical conductivity (Ngodhe et. al., 2014). There are decrease in trends for species richness, total abundance and diversity in water bodies affected by high values for chemical parameters in the water (Prenda and Gallardo-Mayenco 1996; Heino 2000).

Around the world, freshwater habitats are being subjected to increased levels of human disturbance (Saunders et al., 2002). Globally, of all the inland water ecosystems, the most threatened river catchments are found in the Indian subcontinent (World Conservation Monitoring Centre [WCMC], 2000). In general, water beetle communities all around the world suffer from desertification, irrigation, eutrophication induced by livestock, man, and agriculture, as well as overall loss of primary habitats (Jach and Balke, 2008).

Water beetles in Himalayas is not much explored and the distribution of certain dytiscid genera in the Himalayas is still very incomplete (Brancucci and Wewelka, 2005). Bhutan is still regarded as one of the least explored countries in terms of aquatic entomology. Namgaychholing Gewog (block) under Samtse Dzongkhag (district) has a rich freshwater ecosystem encompassing different elevational gradient. Conversely, the study area is expected to have a high diversity of water beetles distributed across different habitats. This study aims to analyse the diversity and distribution of water beetles in different freshwater habitats and their response to physico-chemical parameters. The study will augment the scientific exploration in beetle diversity in Bhutan and especially, it will spearhead to form the baseline data for the water beetles in Bhutan as this is the first of its kind in the country.

International Journal of Environment

ISSN 2091-2854
Materials and methods

Study area

Namgaychholing Gewog (88°57'57.935"E and 27°5'33.421"N) is located in Samtse Dzongkhag to the North-West of Dzongkhag headquarter. It covers around 128.49 square kilometers of area (Dzongkhag Administration Samtse, 2018).

The Gewog experiences annual rainfall ranging from 1200 mm to 3000 mm. The Gewog has hot and humid summer, dry and cool winter with snowfall in the northern part of Gewog and falls in sub-tropical monsoon zone. The Gewog has red, alluvial and black types of soil. Namgaychholing Gewog consist of five Chiwogs (village) (Dzongkhag Administration Samtse, 2018).

Research Design

Sampling

The sampling were carried out following stratified random sampling with equal proportion (Williams et al., 2007). Water beetles were collected from four different habitats (rivers, streams, water logged areas and marshy areas) of the gewog. Ten sample sites were selected from each habitat. Therefore, from four different habitat the total sampling site was 40 from different areas. The sampling unit for each habitat were selected through reconnaissance survey. In each sample site 1 hour was set as a standard time, including time spent sweeping and removing the debris and unwanted insects (Turic et al., 2017). The sample size of 3m × 1 m
was adopted following Turic et al. (2017). Data collection was done during the pre-monsoon season before the water beetles were not affected by the monsoon vagaries.

Data collection
Water net and household sieve of various pore size was used to collect the specimens (Turner, 2007; Boukal et al., 2007). In rivers, streams and water-logged areas, s kicking, brushing and rubbing the surface and then gently sweeping with the water net was done to collect the water beetles. The small water beetles that were seen floating on the surface of the water were collected by gently sweeping the surface with household sieve. The sampling techniques followed Lhundup and Dorji (2018). If the habitats were covered with woody debris and leaf litter, it was disturbed with poles and leg, and was swept with water net. The collected samples were washed thoroughly into the net and the water beetles were picked with forceps. In greater depth of water, it was disturbed with the help of poles. This was done to make sure that the water beetles that were inside the sand and soil are mixed with water and brought on the surface. In marshy areas with minimum water, it was disturbed with the sticks and household sieve was used to collect the specimens.

Measurement of physico-chemical parameters
All three physico-chemical parameters were measured in the field. The pH meter was used to measure water pH and pen type DO₂ meter to measure both water temperature and DO at every sampling unit before sampling.

Fixation, preservation and identification
The specimens were fixed using 10% formalin and preserved in 70% ethanol (Gaston et al., 1996). Specimens which could be identified on the field were recorded, photographed and released back to their natural habitats. The collected samples were identified later in the laboratory using microscope, identification keys, monographs and assistance from taxonomic expert.

Data analysis
The species diversity was analysed using Shannon’s Diversity Index. The taxon richness was computed using \( R = (S-1) / \log N \) (Wilson, 1992) and Evenness using Pielou evenness formula \( E = H' / H_{max} \) (McGinley, 2014). Berger-Parker’s Dominance index (Das and Gupta, 2010) was used to find the dominance in a habitat. Relative abundance (Purkayastha and Gupta, 2012) was calculated for individual family and species. Kruskal Wallis test was done using R studio to compare the variations in diversity, richness and abundance among different habitats. Cluster analysis was done between four different habitats with their relative abundance (RA) using PC-ORD.
The Pearson’s correlation was used to find the relationship between diversity indices and physico-chemical parameters. Furthermore, Canonical correspondence analysis was done using PC-ORD to find the distribution of water beetles according to physico-chemical parameters. Canonical correspondence analysis was done to find the distribution of water beetles according to the physico-chemical parameters

Shannon-Wiener Index ($H'$) (Konopinski, 2020)

$$H' = - \sum_{i=1}^{S} P_i \times \ln (P_i)$$

Where, $P_i = n_i/N$ (RA of each species, or proportion),

$n_i$ = no. of individuals in species $i$

$N$ = total number of all individuals

Berger-Parker Dominance ($P_{max}$)

Dominance, $D = P_{max}$

Where, $n_i$ = number of individuals of a species (of one plot)

$N$ = total number of individuals of all species (of one plot)

Relative Abundance

$$Relative \ Abundance \ (RA) = \frac{No. \ of \ individual}{Total \ no. \ of \ individuals} \times 100\%$$

Results and discussion

Diversity of water beetles

A total of 1219 water beetles belonging to 7 families with 30 species (Table 1) were recorded. Family Dytiscidae ($n=626$, $RA=51$) was most abundant and family Sphaeriidae ($n=3$, $RA=0.25$) the least. Amongst all the species, *Lacconectus basalis* Sharp, 1882 ($n=416$, $RA=34.13$) of Dytiscidae family was most abundant and the least were *Cybister lateralis marginalis* De Geer, 1774 ($n=1$, $RA=0.08$) and *Aculomicrus alesi* Fikacek, 2010 ($n=1$, $RA=0.08$).

*Lacconectus basalis* was most abundant probably due to its adaptability to wide range of physico-chemical parameters (Table 2). The member of Dytiscidae family were the most abundant due to wide range of adaptability in different habitats (Freitag *et al.*, 2016) as compared to other families. The Dytiscidae family was found in all four habitats but Sphaeriidae was found only in streams and waterlogged areas. Dytiscidae family was also found to be the most abundant in Loktak Lake of Manipur, India (Devi *et al.*, 2014).

Overall, Shannon diversity index was $H' = 2.48$, evenness $E_H = 0.73$, taxon richness $S_R = 4.08$ and dominance $P_{max} = 0.34$. Among the different habitats, marshy area had the highest Shannon diversity index.


\( H' = 2.45 \), evenness \( (E_H = 0.75) \), taxon richness \( (S_R = 2.03) \) and dominance \( (P_{imax} = 0.33) \) followed by river with diversity \( (H' = 2.00) \), evenness \( (E_H = 0.70) \), taxon richness \( (S_R = 2.78) \) and dominance \( (P_{imax} = 0.41) \). Waterlogged area had diversity \( (H' = 1.92) \), evenness \( (E_H = 0.75) \), taxon richness \( (S_R = 2.02) \) and dominance \( (P_{imax} = 0.39) \) and the least was found in streams with diversity \( (H' = 1.82) \), evenness \( (E_H = 0.69) \), taxon richness \( (S_R = 2.38) \) and dominance \( (P_{imax} = 0.275) \) (Figure 3). The difference in these results among different habitats could be because, the diversity of macro invertebrates is directly influenced by the surrounding land use, which in turn influence water pH, temperature, salinity and DO (Dhakal, 2006).

Figure 2: Shannon Wiener Diversity, Evenness and Richness of water beetles in different habitats.

A Kruskal Wallis test showed no significant differences in water beetle diversity \( (H(3) = 1.901, p = .593) \), richness \( (H(3) = 1.811, p = .612) \), and abundance \( (H(3) = .279, p = .964) \) among different habitats. Gimenez et al. (2015) also found no significant differences between urban and rural streams for the diversity and abundance of species. It was because the physical and chemical properties of the water were different between urban and rural streams. Consequently, the insect community responded to these differences and the species composition, diversity and abundance differed between streams.

Similarly, no significant difference among different habitats in current study maybe due to different habitat with different physical and chemical properties. This chemical and physical properties influenced Beetles community to respond to these differences which resulted in no significant differences among the habitat.
Table 1: Water beetles in the study area with family, species, number of individuals and relative abundance (RA).

| Family       | Species                              | Count | RA (Family) | RA (Species) |
|--------------|--------------------------------------|-------|-------------|--------------|
| Dytiscidae   | *Copelatus sibelaemontis* Balke, 2010| 62    | 51          | 5.09         |
|              | *Cybister lateralimarginalis* De Geer, 1774 | 1     | 0.08        |              |
|              | *Lacconectus basalis* Sharp, 1882   | 416   |             | 34.13        |
|              | *Laccophilus chinensis* Boheman, 1858| 129   |             | 10.58        |
|              | *Hydroglyphus geminus* (Fabricius, 1792) | 7     |             | 0.57         |
|              | *Platynectes kashmiranus* Balfour-Browne, 1944 | 11   |             | 0.90         |
| Elmidae      | *Stenelmis sandersoni* Musgrave, 1940 | 25    | 2.05        | 2.05         |
| Gyrinidae    | *Patrus palawanensis* Regimbart, 1907 | 91    | 16.82       | 7.47         |
|              | *Dineutus spinosus* Fabricius, 1781  | 56    |             | 4.59         |
|              | *Dineutus unidentatus* Aube, 1838    | 14    |             | 1.15         |
|              | *Metagyrinus arrowi* (Regimbart, 1907) | 44   |             | 3.61         |
| Hydraenidae  | *Hydraena rufipes* Curtis, 1830      | 60    | 9.68        | 4.92         |
|              | *Limnebius rubropiceus* Kuwert, 1890 | 9     |             | 0.74         |
|              | *Limnebius truncatellus* (Thunberg, 1794) | 46   |             | 3.77         |
|              | *Ochthebius opacipennis* Champion, 1920 | 3     |             | 0.25         |
| Hydrophilidae| *Aculomicrus alesi* Fikacek, 2010    | 1     | 16.08       | 0.08         |
|              | *Agraphydrus flavonotus* (Komarek, 2018) | 2     |             | 0.16         |
|              | *Agraphydrus indicus* (Orchymont, 1932) | 15    |             | 1.23         |
|              | *Agraphydrus kempi* (Orchymont, 1922) | 26    |             | 2.13         |
|              | *Dactylosternum circuncinctum* Fikacek, 2010 | 2    |             | 0.16         |
|              | *Enochrus esuriens* (Walker, 1858)   | 80    |             | 6.56         |
|              | *Enochrus subsignatus* (Harold, 1877) | 9     |             | 0.74         |
|              | *Helochaera atropiceus* Regimbart, 1903 | 22   |             | 1.80         |
Species composition of water beetles

The 7 families of water beetles found in the study area were Hydrophilidae, Dytiscidae, Gyrinidae, Hydraenidae, Noteridae, Sphaeriusidae and Elmidae. The family Hydrophilidae was found to have 13 different species making 43.3% (Table 1) of total species, followed by Dytiscidae with 6 species making 20%. The family Gyrinidae and Hydraenidae consist of 4 species contributing 14% by each family and the least was Noteridae, Sphaeriusidae and Elmidae with 1 species each making 3% of total species.

Though Dytiscidae was found abundantly in terms of number of species composition, Hydrophilidae had the high number of species. However, Dytiscidae had higher species composition than Hydrophilidae in Coleopteran Water Beetles in Kenyir Water Catchment of Terengganu, Malaysia (Abdullah, 2009). However, Freitag et al. (2016) reported that species composition was higher in family hydrophilidae than in Dytiscidae. The difference in findings could be because Abdullah (2009) studied in single habitat (Lake) but Freitag et al. (2016) did from both aquatic and riparian habitats. Similarly, this study area covers four different habits, hence species composition maybe higher in family Hydrophilidae than in Dytiscidae.

Distribution of water beetles in different habitats

The families Hydrophilidae, Dytiscidae, Gyrinidae and Hydraenidae was distributed in all four habitats (Figure 4). The species belonging to Noteridae ($n=1$) were found each in waterlogged and in marshy habitat. In family Elmidae, species ($n=1$) was found only in marshy habitat and Sphaeriusidae ($n=1$) were encountered each in streams and waterlogged habitat but was of same species. The highest number of species ($n=8$) of Hydrophilidae and ($n=5$) of Dytiscidae were encountered in riverine habitat and lowest ($n=5$) each in waterlogged and marshy habitats, and ($n=2$) in streams respectively. However, highest number of species for Gyrinidae ($n=3$) were recorded from streams and lowest number ($n=2$) each in other three habitats. In
family Noteridae \((n=1)\) species was found each in waterlogged and marshy habitats. In Elmidae and Sphaeriusidae \((n=1)\) species was distributed in waterlogged habitat and streams respectively (Figure 6). The distribution of individual species with physico-chemical parameters in four different habitats is shown in Table 2. However, in some species there are only one value and no range for parameters because those species were found only in one plot.

The cluster analysis based on the relative abundance of taxa grouped four different habitats into three major clades. The dendrogram showed that the habitat river and waterlogged area have similar species composition having closer, 100% similarity in dendrogram scale (Figure 4). The dominant species were *Lacconectus basalis*, *Limnebius truncatellus*, *Copelatus sibelaemontis*, *Enochrus esuriens* and *Laccophilus chinensis*. The similarity among these habitats could be because the plots in habitat rivers were sampled in riverine habitats, which moreover, are similar to waterlogged areas with similar physiochemical parameters.

![Figure 3: Cluster analysis of different habitats using dendrogram based on relative abundance.](image)

Similarly, river, waterlogged area and stream habitat were having 50% similarity (Figure 4). These indicate that these three habitats have similar water physiochemical parameters such as pH, water temperature and dissolved oxygen, which made the aquatic habitat suitable for that particular species of water beetle (Dorji, 2014).

The family Hydrophilidae, Dytiscidae, Gyrinidae and Hydraenidae are adapted to variety of freshwater habitats. The adults and larvae of most species of family Hydrophilidae lives in stagnant water, running water, seepages, numerous species are reportedly riparian or terrestrial (Jach and Balke, 2008). The species of family Dytiscidae inhibits stagnant water, running water, groundwater, seepages and phytotelmata (Khan and Ghosh, 2001). The family Gyrinidae are found commonly in fresh water ponds, lakes, open flowing streams and majority of the species are found in running water (Sharma et al., 2019). The adults of Hydraenidae are aquatic and lives in stagnant water, running water, seepages any are riparian (Jach and Balke, 2008). The family Noteridae are commonly found in stagnant water between roots of water plants (Nilsson and Van, 2005). The most species of Elmidae occurs in well-aerated streams and rivers but some can be found in shores of lakes (Elliott, 2008). The Sphaeriusidae beetles are found in accumulations of gravel, sand and mud at the edges of streams and rivers (Lawrence and Slipinski, 2013).
**Figure 4:** Distribution of water beetle families in different habitats.

**Table 2:** Distribution of water beetle species with physico-chemical parameters in four habitats.

| Species                        | Physico-chemical parameters | Habitats |
|--------------------------------|------------------------------|----------|
|                                | pH     | Temp (°C) | DO (ppm) | Stream | River | WA | MA |
| **Copelatus sibelaemontis**    | 9.1 - 10.08 | 13.6 - 18.1 | 0.73 - 6.34 | - | - | + | + |
| **Cybister lateralinmarginalis** | 9.5    | 19.5 | 6.43 | - | + | - | - |
| **Lacconectus basalis**        | 7.75 - 10.48 | 12.7 - 21.1 | 0.73 - 6.98 | + | + | + | + |
| **Laccophilus chinensis**      | 8.41 - 9.96 | 19.5 - 25.1 | 2.29 - 6.43 | + | + | + | + |
| **Hydroglyphus geminus**       | 8.98   | 26.5 | 5.66 | - | + | - | - |
| **Platynectes kashmiranus**    | 9.07 - 9.77 | 14.7 - 19.5 | 4.39 - 6.43 | - | + | - | + |
| **Stenelmis sandersoni**       | 9.78   | 12.7 | 5.61 | - | - | - | + |
| **Patrus palawanensis**        | 8.87 - 11 | 12.7 - 22.8 | 5.03 - 6.29 | + | + | + | + |
| **Dineutus spinosus**          | 8.57 - 11 | 14.2 - 23.3 | 2.45 - 6.19 | + | - | + | + |
| **Dineutus unidentatus**       | 8.94 - 9.7 | 19.6 - 21.1 | 5.42 - 6.1 | - | + | - | - |
| **Metagyrinus arrowi**         | 8.87 - 9.96 | 12.7 - 21.1 | 5.03 - 6.29 | + | - | - | + |
| **Hydraena rufipes**           | 9.07 - 9.84 | 12.7 - 18.1 | 0.73 - 6.29 | + | - | + | + |
| **Limnebius rubropiceus**      | 9.77   | 16.9 | 4.39 | - | - | - | + |
**Limnebius truncatellus**  9.5 - 9.74  15.5 - 20.8  5.33 - 6.1  +  +  -  -  

**Ochthebius opacipennis**  9.7  19.6  6.1  -  +  -  -  

**Aculomicrus alesi**  10.48  14.2  5.82  +  -  -  -  

**Agraphydrus flavonotus**  9.98  13.6  6.34  -  -  -  +  

**Agraphydrus indicus**  9.5  19.5  6.43  -  +  -  -  

**Agraphydrus kempi**  9.09 - 10.24  14.1 - 22.6  0.73 - 5.23  -  +  +  -  

**Dactylosternum circumcinctum**  10.48  14.2  5.82  +  -  -  -  

**Enochrus esuriens**  8.87 - 10.06  12.7 - 21.1  0.73 - 6.98  +  +  +  +  

**Enochrus subsignatus**  9.09 - 9.31  19.6 - 22.6  5.02 - 6.1  -  +  +  -  

**Helochares atropiceus**  8.87 - 9.77  16.9 - 19.5  4.39 - 6.43  +  +  -  +  

**Helochares lentus**  8.98  26.5  5.66  -  +  -  -  

**Helochares taprobanicus**  9.9 - 9.96  19.6 - 21.1  5.03 - 5.5  +  -  -  +  

**Laccobius celsus**  9.31 - 10.45  12.7 - 24.8  4.95 - 6.34  +  -  +  +  

**Laccobius patruelis**  9.5 - 10.24  14.1 - 19.5  4.82 - 6.43  -  +  +  -  

**Laccobius simulans**  8.98  26.5  5.66  -  +  -  -  

**Mesonoterus laevicollis**  9.1 - 10.45  14.1 - 16.9  0.73 - 4.95  -  -  +  +  

**Sphaerius minutus**  9.96 - 10.24  18.1 - 21.1  5.03 - 5.23  +  -  +  -  

[Temp = Temperature, DO = Dissolved oxygen, WA = Waterlogged area, MA = Marshy area]  
[+ Recorded, - Not recorded]

**Relationship between water beetles and physico-chemical parameters**

The highest mean for pH (9.93) was found in stream and lowest (9.43) in marshy areas. The mean water temperature for riverine (19.77) was highest and minimum mean (16.39) was recorded in streams, and the highest mean (5.81) for dissolved oxygen was recorded in riverine and lowest (4.45) in waterlogged areas (Table 3).
Table 3: Mean, standard error and (minimum and maximum values) of physico-chemical parameters of different habitats.

| Habitat              | pH                   | Temperature (°C)          | Dissolved oxygen (ppm) |
|----------------------|----------------------|---------------------------|------------------------|
| Stream               | 9.93 ± 0.19 (8.87-11.00) | 16.39 ± 0.93 (12.70-21.10) | 5.77 ± 0.15 (5.03-6.37) |
| River                | 9.50 ± 0.14 (8.94-10.42) | 19.77 ± 1.09 (14.60-26.50) | 5.81 ± 0.19 (5.08-6.98) |
| Waterlogged area     | 9.50 ± 0.29 (7.75-10.48) | 18.20 ± 1.33 (13.50-25.10) | 4.45 ± 0.51 (0.73-5.91) |
| Marshy area          | 9.43 ± 0.14 (8.57-9.98)  | 17.86 ± 1.23 (12.70-24.80) | 5.25 ± 0.36 (2.45-6.34) |

A Pearson correlation test showed no significant association between diversity and physico-chemical parameters with pH ($r_s = .217, p =.178$), temperature ($r_s = -.297, p = .063$) and dissolved oxygen ($r_s = -.009, p = .956$). There was also no significant correlation between taxon richness and pH ($r_s = .269, p = .093$), dissolved oxygen ($r_s = .031, p = .848$) and with temperature ($r_s = -.261, p = .103$). The evenness with pH showed a significant moderate negative correlation ($r_s = -.463, p = .003$) and weak positive correlation with temperature ($r_s = -.370, p = .019$) but no significant correlation with dissolved oxygen ($r_s = -.046, p = .778$). There was significant weak negative correlation between abundance and physico-chemical parameters with dissolved oxygen ($r_s = -.330, p = .038$) but indicated no significant correlation with temperature ($r_s = -.237, p = .141$) and pH ($r_s = .107, p = .512$).

A pH range of 6.5 to 9.0 is the optimal for the life of freshwater bottom-dwelling macroinvertebrates (USEPA, 1986). There is a positive correlation between species richness and pH of aquatic insects (Dalal and Gupta, 2016). Similarly, the taxa richness of invertebrates and diversity increases with increases in pH (Scheibler et al., 2014; Flores and Zafaralla, 2012).

The negative correlation between diversity and richness with water temperature was recorded could mean that water beetles are less tolerant to warmer water. It is because when temperature increases dissolved oxygen and pH decreases making less favorable condition for water beetles. This is due to suspended particles in the water which absorb the heat resulting to increase the water temperature (Paaijmans et al., 2008). When temperature increases, the dissolved oxygen in the water decreases because warm water holds less dissolved oxygen than the cold water (Mandal, 2014). According to Ngodhe et al. (2014), temperature negatively influences the species diversity, dominance and richness of a macroinvertebrates.

The dissolved oxygen was negatively correlated with diversity, evenness and abundance but was positively correlated with taxon richness in the study area. The concentration of dissolved oxygen ($\infty$) is one of the most important physico-chemical parameters to determine water quality and to know the distribution of different aquatic insects (Wahizatul et al., 2011). However, a study on aquatic insect biodiversity and water
quality parameters of streams in northern Thailand found that dissolved oxygen negatively influences the taxa richness (Prommi and Payakka, 2015).

**Distribution of water beetles in relation to physico-chemical parameters**

Canonical correspondence analysis used to find the distribution of water beetles in relation to the physico-chemical parameters (Figure 5) where the eigenvalues for axis 1 was 0.498, axis 2 was 0.247, and axis 3 was 0.183. The highest eigenvalue was found in axis 1 and the variable corresponded to temperature. Dissolved oxygen ($r = -0.605$) was negatively correlated with axis 2 and pH ($r = 0.430$) was positively correlated with Axis 3. Temperature ($r = 0.760$) was positively correlated with Axis 1. Temperature was having the highest correlation with Axis 1 corroborating to be the main determining factor in distribution of water beetle. Water temperature is the most effective environmental variable in development rate in terrestrial and aquatic ectotherms and insects are affected from the changes of temperature (Ragland and Kingsolver, 2008; Chuche and Thiery, 2012).

![Figure 5: Distribution of water beetle families in different habitats.](image)

Distribution of *Dineutus unidentatus*, *Agraphydrus kempi*, *Laccophilus chinensis*, and *Dineutus spinosus* were positively related with temperature. For distribution of species like *Dactylosternum circumcinctum*, *Aculomicrus alesi*, *Enochrus esuriens* and *Agraphydrus flavonotus* the parameters pH and dissolved oxygen are important factors as the direction of arrow is parallel towards the negative of axis 2, however, dissolved oxygen had more correlation compared to pH (Garca-Criado et al., 1999).
Limitation

The study was conducted during pre-monsoon season only. All season data collection would yield more diversity of water beetles and help in better comparison of diversity, composition and distribution among different seasons of the year. Albeit there are many physico-chemical parameters that affects the aquatic environment, the unavailability of instrument for all the parameters had posed constraint to study only current parameters. Consideration of other physico-chemical parameters affecting the water beetles in the freshwater ecosystem would shape the better understanding of the variation of diversity, composition and distribution of water beetles in different habitats.

Conclusion

A total of 1219 water beetles belonging to 7 different families with 30 species were recorded from the study area. The family Dytiscidae was the most common and the family Sphaeriusidae was the least in the whole study area. The whole study area had Shannon diversity index of 2.48, while comparing in different habitats, marshy area had highest Shannon diversity index and lowest in streams. A Kruskal Wallis test on water beetle diversity among different habitats indicated no significant differences in water beetle diversity. It was because the sampling was conducted in four different habitat that have different physical and chemical properties. The families Hydrophilidae, Dytiscidae, Gyrinidae and Hydraenidae was distributed in all four habitats. The dendrogram showed that the habitat river and waterlogged area have similar species composition having very closer, 100% similarity in dendrogram scale.

Pearson’s correlation indicated no significant correlation between diversity and taxon richness with all three physico-chemical parameters. The evenness with pH showed moderate negative correlation and weak positive correlation with temperature. Canonical correspondence analysis found temperature as a main determining factor in distribution of water beetle.

With broad families encompassing the water beetles, similar study on water beetles in different regions of the country is recommended with particular emphasis on different families. Moreover, Bhutan falls in junction of Palearctic and Oriental Zoogeographic regions, which are known to be rich in diversity of water beetles. Hence, study in high-altitude lakes, which are assumed to be rich in diversity of water beetles are recommended despite social beliefs to keep lakes isolated and undisturbed.

Conflict of interest

The authors declare that they have no affiliations or involvement in any organization or entity with any financial or non-financial interests in the matter to the work reported in this paper.
Author contribution statements

Mr. Tez Bdr. Ghalley conceived and designed the analysis. He was also actively involved in collection of data from the field. In addition, analyzed the data and wrote the first draft of manuscript.

Mr. Ugyen Dorji helped supervise the study. He helped in developing research framework and proposal since the initial stage. As a corresponding author, he also performed the analysis using statistical tools R and reviewed the manuscript after getting comments from reviewers.

Mr. Cheten Dorji with his knowledge in entomology helped in identification of specimens. He also gave critical inputs in formulating research framework and reviewed the manuscript.

Mr. Arjun Nepal and Mr. Namgay Shacha contributed to the interpretation of the results and provided critical feedback and helped shape the research, analysis and manuscript. Both were also involved in planning and supervising the field works.

Acknowledgement

Authors would like to thank Mr. Drukpola, Officiating Program Director of National Research & Development Centre for Aquaculture, Gelephu, Bhutan for providing instruments and College of Natural Resources for their support in successful completion of this paper.

References

Abdullah, F., 2009. Diversity and New Records of Coleopteran Water Beetles (Dytiscidae, Hydrophilidae) in Kenyir Water Catchment of Terengganu, Malaysia. *International Journal of Zoological Research*, 5 (1), 1-8. DOI: http://dx.doi.org/10.3923/ijzr.2009.1.8

Boukal, D. S., Sabelis, M. W. & Berec, L., 2007. How predator functional responses and Allee effects in prey affect the paradox of enrichment and population collapses. *Theoretical Population Biology*, 72(1), 136-147. DOI: https://doi.org/10.1016/j.tpb.2006.12.003

Chuche, J. & Thiery, D., 2012. Egg incubation temperature differently affects female and male hatching dynamics and larval fitness in a leafhopper. *Ecology and evolution*, 2(4), 732-739. DOI: https://doi.org/10.1002/ece3.89

Dalal, A. & Gupta, S., 2016. A comparative study of the aquatic insect diversity of two ponds located in Cachar District, Assam, India. *Turkish Journal of Zoology*, 40(3), 392-401. DOI: https://doi.org/10.3906/zoo-1505-18

Das, K., & Gupta, S., 2010. Aquatic Hemiptera community of agricultural fields and rain pools in Cachar District, Assam, North East India. *Assam University Journal of Science and Technology*, 5(1), 123-128. https://www.entomoljournal.com/vol3Issue1/pdf/3-1-17.1.pdf (accessed on: 19 April, 2019).
Devi, M. B., Devi, O. S. & Singh, S. D., 2014. Diversity, Abundance and Species Composition of Water Beetles (Coleoptera: Dytiscidae, Hydrophilidae and Noteridae) from the Loktak Lake of Manipur, North East India. *World Journal of Zoology*, 9(1), 04-12. DOI: https://doi.org/10.5829/idosi.wjz.2014.9.1.82213

Dhakal, S., 2006. Study on Physiochemical Parameters and Benthic Macroinvertebrates of Balkhu Khola in Kathmandu Valley, Central Nepal. *Management of Water, Wastewater and Environment: Challenges for the Developing Countries, Kathmandu*. https://lib.icimod.org/api/files/90160a3c-c89f-4dca-b057-8b55e9e15b28/630.pdf (accessed on: 1 May, 2019).

Dong, B., Geng, C., Cai, Y. & Ji, L., 2014. Aquatic Coleoptera response to environmental factors of freshwater ecosystems in Changbai Mountain, northeast China. *Aquatic ecosystem health & management*, 17(2), 171-178. DOI: http://dx.doi.org/10.1080/14634988.2014.910441

Dorji, T., 2014. *Macroinvertebrates diversity in response to environmental variables in headwater streams*. Project Report. Royal University of Bhutan: Thimphu.

Dzongkhag Administration Samtse (2018) Royal Government of Bhutan. Dzongkhag Administration Samtse. http://www.samtse.gov.bt/gewogs/namgyacholinglahireni-gewog (accessed on: 15 August, 2018).

Elliott, J. M., 2008. The ecology of riffle beetles (Coleoptera: Elmidae. *Freshwater Reviews*, 1(2), 189-204. https://fba.org.uk/common/Uploaded%20files/Journals/20081105ElliottNS.pdf. (accessed on: 23 June, 2019).

Epler, J. H., 1996. Identification manual for the water beetles of Florida. *Department of Environmental Protection, Division of Water Facilities. Tallahassee, Florida*. 259p. http://johnepler.com/FLWB96.pdf (accessed on: 23 June, 2019).

Flores, M. J. L. & Zafaralla, M. T., 2012. Macroinvertebrate composition, diversity and richness in relation to the water quality status of Mananga River, Cebu, Philippines. *Philippine Science Letters*, 5(2), 103-113. https://philsciletters.net/2012/2012n2.11p12.pdf (accessed on: 19 March, 2019).

Freitag, H., Jach, M. A. & Wewalka, G., 2016. Diversity of aquatic and riparian Coleoptera of the Philippines: checklist, state of knowledge, priorities for future research and conservation. *Aquatic insects*, 37(3), 177-213. DOI: https://doi.org/10.1080/01650424.2016.1210814

Garca-Criado, F., Fernandez-Alaez, C. & Fernandez-Alaez, M., 1999. Environmental variables influencing the distribution of Hydraenidae and Elmidae assemblages (Coleoptera) in a moderately-polluted river basin in north-western Spain. *European Journal of Entomology*, 96, 37-44. https://www.eje.cz/pdfs/eje/1999/01/08.pdf (accessed on: 10 March, 2019).
Gerdes, G., Spira, J. & Dimentman, C., 1985. The fauna of the Gavish Sabkha and the Solar Lake—a comparative study. In Hypersaline Ecosystems Springer, Berlin, Heidelberg, 322-345 DOI: https://doi.org/10.1007/978-3-642-70290-7_18

Ghannem, S., Touaylia, S. & Boumaiza, M., 2018. Beetles (Insecta: Coleoptera) as bioindicators of the assessment of environmental pollution. Human and Ecological Risk Assessment: An International Journal, 24(2), 456-464. DOI: https://doi.org/10.1080/10807039.2017.1385387

Gimenez, B. C. G., Lansac-Toha, F. A. & Higuti, J., 2015. Effect of land use on the composition, diversity and abundance of insects drifting in neotropical streams. Brazilian Journal of Biology, 75(4), 52-59. DOI: https://doi.org/10.1590/1519-6984.03914

Heino, J., 2000. Lentic macroinvertebrate assemblage structure along gradients in spatial heterogeneity, habitat size and water chemistry. Hydrobiologia, 418(1), 229-242. DOI: https://doi.org/10.1023/A:1003969217686

Khan, D. R. & Ghosh, L. K., 2001. Faunal diversity of aquatic insects in freshwater wetlands of South Eastern West Bengal. Zoological Survey of India. http://faunaofindia.nic.in/PDFVolumes/occpapers/194/index.pdf (accessed on: 10 March, 2019).

Konopiński, M. K., 2020. Shannon diversity index: A call to replace the original Shannon’s formula with unbiased estimator in the population genetics studies. PeerJ, 2020(6). DOI: https://doi.org/10.7717/peerj.9391

Lawrence, J. & Slipinski, A., 2013. Australian beetles volume 1: morphology, classification and keys . Csiro publishing. 171 (1), 226-226. DOI: https://doi.org/10.1111/zoi.12135

Lhundup, K. & Dorji, U., 2018. Macro-invertebrate Diversity and its relationship with environmental variables in Adha Lake between monsoon and post-monsoon seasons. Bhutan Journal of Natural Resources & Development. 5 (1): 13-24. DOI: https://doi.org/10.17102/cnr.2018.02

Liang, Z. L. & Jia, F., 2018. A new species of Sphaerius Waltl from China (Coleoptera, Myxophaga, Sphaeriusidae. Zoo Keys, 808, 115-121. DOI: https://dx.doi.org/10.3897/zookeys.808.30600

Mandal, H. K., 2014. Assessment of wastewater temperature and its relationship with turbidity. Recent Research in Science and Technology, 6(1), 258-262. http://recent-science.com/index.php/rrst/article/view/19081/9567 (accessed on: 20 October, 2018)

McGinley, M., 2014. Species richness. http://www.eoearth.org (accessed on: 1 December 2018).

Ngodhe, S. O., Raburu, P. O. & Achieng, A., 2014. The impact of water quality on species diversity and richness of macroinvertebrates in small water bodies in Lake Victoria Basin, Kenya. Journal of Ecology and the Natural Environment, 6(1), 32-41. DOI: https://doi.org/10.5897/JENE2013.0403
Nilsson, A. N. & Van Vondel, B. J., 2005. *Amphizoidae, Aspidytidae, Haliplidae, Noteridae and Paelobiidae (Coleoptera, Adephaga)*. BRILL. DOI: https://doi.org/10.14411/eje.2006.070

Paaijmans, K. P., Takken, W., Githeko, A. K. & Jacobs, A. F. G., 2008. The effect of water turbidity on the near-surface water temperature of larval habitats of the malaria mosquito Anopheles gambiae. *International journal of biometeorology*, 52(8), 747-753. DOI: https://doi.org/10.1007/s00484-008-0167-2

Prenda, J. & Gallardo-Mayenco, A., 1996. Self-purification, temporal variability and the macroinvertebrate community in small lowland Mediterranean streams receiving crude domestic sewage effluents. *Archiv fur Hydrobiologie. Stuttgart*, 136(2), 159-170. DOI: https://doi.org/10.1127/archiv-hydrobiol/136/1996/159

Prommi, T. & Payakka, A., 2015. Aquatic insect biodiversity and water quality parameters of streams in northern Thailand. *Sains Malaysiana*, 44(5), 707-717. DOI: http://dx.doi.org/10.17576/jsm-2015-4405-10

Purkayastha, P., & Gupta, S., 2012. Insect diversity and water quality parameters of two ponds of Chatla Wetland, Barak Valley, Assam. *Current World Environment*, 7(2), 243. DOI: http://dx.doi.org/10.12944/CWE.7.2.08

Ragland, G. J. & Kingsolver, J. G., 2008. The effect of fluctuating temperatures on ectotherm life-history traits: comparisons among geographic populations of Wyeomyia smithii. *Evolutionary Ecology Research*, 10(1), 29-44. https://jgking.web.unc.edu/wp-content/uploads/sites/1829/2012/06/Ragland_Kingsolver_EER_2008.pdf (accessed on: 24 June 2019).

Saunders, D. L., Meeuwig, J. J. & Vincent, A. C. J., 2002. Freshwater protected areas: strategies for conservation. *Conservation Biology*, 16(1), 30-41. DOI: https://doi.org/10.1046/j.1523-1739.2002.99562.x

Scheibler, E. E., Claps, M. C. & Roig, S. A., 2014. Temporal and altitudinal variations in benthic macroinvertebrate assemblages in an Andean river basin of Argentina. DOI: https://doi.org/10.4081/jlimnol.2014.789

Sharma, S., Sharma, G. & Pir, F. A., 2019. Diversity and habitat selection of aquatic beetles (Coleoptera. *Journal Of Pharmacy And Biological Sciences*, 14(1), 31-37. http://www.iosrjournals.org/iosr-jpbs/papers/Vol14-issue1/Version-1/E1401013137.pdf (accessed on: 24 June 2019).

Turic, N., Temunovic, M., Vignjevic, G., Dunic, J. A., & Merdic, E., 2017. A comparison of methods for sampling aquatic insects (Heteroptera and Coleoptera) of different body sizes, in different habitats.
using different baits. European Journal of Entomology, 114, 123-132. DOI: http://dx.doi.org/10.14411/eje.2017.017

Turner, R. C., 2007. Chapter 23. Collecting water beetles: an introduction. Guides to the Freshwater Invertebrates of Southern Africa Vol. 10 Aquatic Coleoptera, 193-204. https://dep.wv.gov/WWE/getinvolved/sos/Documents/Benthic/UMW/Coleoptera.pdf (accessed on: 24 June 2019).

United States Environmental Protection Agency., 1986. Quality Criteria for Water. Office of Water Regulations and Standards. Washington, DC EPA/440/5-86-001. https://www.epa.gov/sites/production/files/2018-10/documents/quality-criteria-water-1986.pdf (accessed on: 24 June 2019).

Wahizatul, A. A., Long, S. H. & Ahmad, A., 2011. Composition and distribution of aquatic insect communities in relation to water quality in two freshwater streams of Hulu Terengganu, Terengganu. Journal of Sustainability Science and Management, 6(1), 148-155. academia.edu/4233611/Composition_and_distribution_of_aquatic_insect_communities_in_relation_to_water_quality (accessed on: 24 June 2019).

Wilson, E. O., 1992. The Diversity of Life. New York, W. NY: WW Norton & Company, 48. https://www.goodreads.com/book/show/503051.The_Diversity_of_Life (accessed on: 24 June 2019).

Williams, P., Whitfield, M., & Biggs, J., 2007. How can we make new ponds biodiverse? A case study monitored over 7 years. In Pond Conservation in Europe. Springer, Dordrecht. 137-148. DOI: http://dx.doi.org/10.1007/978-90-481-9088-1_12

World Conservation Monitoring Centre., 2000. Global Biodiversity: Earth’s living resources in the 21st century. By: Groom bridge, B. and Jenkins, M.D. World Conservation Press, Cambridge, UK. https://www.cabdirect.org/cabdirect/abstract/20026791684 (accessed on: 24 June 2019).