Distribution and Assessment of Heavy Metal Levels Using Geoaccumulation Index and Pollution Load Index in Lake Edku Sediments, Egypt

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Abstract: The aim of this study is to determine and assessment of some heavy metals content (Fe, Mn, Zn, Cu, Ni, Cr, Co and Pb) in the sediment of Lake Edku. Surface sediment samples were collected in September (summer) 2016. The metals content in the Lake sediments decreased in the order of Fe > Mn > Cr > Co > Zn > Cu > Ni = Pb. The range and average ±SD concentrations (µg g⁻¹) of the investigated heavy metals were 30275-43312 (38822±4312) for Fe; 781.8-3432 (1923.6±855) for Mn; 53.8-107 (82.5 ± 18) for Zn; 55.2-91.5 (72.3 ± 13) for Cu; 87.3-158.8 (113.1 ± 24) for Cr; 23.0-63.3 (45.0 ± 13.0) for Ni; and 81.5-117.8 (98.9±11.0) for Co and 32.9 – 55.2 (44.6 ± 8.0) for Pb. The enrichment factor of the study area (EF mean values) have the order of EF₃₅ > EF₃₄ > EF₃₉ > EF₃₈ > EF₃₇ > EF₃₆. According to the geoaccumulation index (Igeo classification), the study area may be practically unpolluted with Fe, Zn, Ni and Cr (Igeo ranged from -2.15 to -0.41) along the study area. On the other hand, the Igeo of Co ranged from moderately to strongly polluted area. In addition, a lower degree of pollution was found in the sediments by the other heavy metals; Pb and Cu (unpolluted to moderate). The Pollution Load Index (PLI) indicated that station IX was characterized by low level of PLI with value of 1.25 while the other stations (ranged from 1.50 to 1.67).

Keywords: Sediment, Heavy Metals, Lake Edku, Enrichment Factor

1. Introduction

Lake Edku lies in the north of the Nile Delta, west of the Rosetta branch between Long. 30° 8\' 30\" & 30° 23\' 00\" E and Lat. 31° 10\' & 31° 18\' N. It is one of four coastal deltaic lakes that are connected to the Mediterranean Sea. Its area has decreased from 28.5x10³ to about 12x10³ Feddans (is an area unit equals 4200sq. meters) as a result of agricultural reclamation. The lake can be divided into three ill-defined basins; eastern, central and western. Lake Edku receives huge amounts of drainage water from three main drains, namely Berzik, Edku and El-Boussili, which open into the eastern basin of the lake [1]. The maximum inflow from all drains is recorded during summer, while the minimum is in winter. An amount of 3.3 x 10⁶ m³ per day of brackish water is introduced into Abu Qir Bay from Lake Edku through Boughaz El-Maadiya [1, 3, 4, 5, 6, 7, 8, 9] have studied the hydrographic and chemical characteristics of Lake Edku water and sediments. Shata & Okbah (2001) [10] studied the geochemical behavior of some trace elements of deep subsurface sediments of Lake Edku. Shata & Okbah (2000) [11] observed the transport of carbonate granules to the subsurface sediments of Lake Edku. Shata (2000) [11] indicated that station IX was characterized by low level of PLI with value of 1.25 while the other stations (ranged from 1.50 to 1.67). The present investigation aims to study the regional distribution of trace metals contamination in the surface
sediment of Lake Edku and to evaluate pollution load index (PLI), enrichment factor (EF) as well as geoaccumulation index (Igeo) of the Lake sediments.

2. Material and Methods

2.1. Study Area and Sampling Sites

Ten stations were selected to cover the whole area of Lake Edku. Surface sediments samples were collected during September (summer) 2016. The sediment samples were collected once from all the selected stations. The positions of these stations are shown in Fig. 1.

2.2. Analytical Methods of Sediments Analysis

The sampling has been carried out by means of plastic tube (PVC), digging gently into the uppermost layer of the sediment. Samples from each site were divided into two sub-samples, then homogenized by mixing and kept in clean plastic containers. Samples were stored frozen until analysis. A sub-sample was taken to determine chemical and physical characteristics of the sediments such as grain size distribution, total organic matter and calcium carbonate contents.

2.2.1. Grain Size Analysis and Total Organic Carbon (TOC)

About 25gm of dried samples was taken for mechanical analysis. The samples were subjected to the combined technique of dry sieving and pipette analysis [12]. Total organic carbon (TOC) was determined according to the method described by Gaudette and Flight [13]. This method utilizes exothermic heating and oxidation with potassium dichromate ($\text{K}_2\text{Cr}_2\text{O}_7$) and concentrated sulphuric acid ($\text{H}_2\text{SO}_4$) of powdered sediment sample and the titration of the excess dichromate with ferrous ammonium sulphate solution using diphenylamine as indicator. The organic carbon is converted to organic matter by multiplying the organic carbon values by the factor of 1.8.

2.2.2. Total Calcium Carbonate ($\text{CaCO}_3$)

The total carbonate was determined by titration technique described by Molina [14]. About 1 g of each sample was treated with 50 ml standardized 0.5N HCl and then released amount of CO$_2$ was determined by back titration with previously standardized 0.25N NaOH solution using phenolphthalin indicator. The percentage of calcium carbonate (% $\text{CaCO}_3$) was determined by the following equation:

\[
\text{CaCO}_3\text{%} = 100 \times 0.05 \times \left\{ \text{NHCl} \times V_{\text{HCl}} - \left( \text{NNaOH} \times V_{\text{NaOH}} \right) \right\}
\]

$\text{NHCl}$: Normality of standardized HCl, $V_{\text{HCl}}$: volume of standardized HCl

$\text{N}_{\text{NaOH}}$: Normality of standardized NaOH, $V_{\text{NaOH}}$: volume of standardized NaOH

2.2.3. Heavy Metal Analysis

The total concentrations of heavy metals (Cu, Cd, Fe, Pb, Zn and Mn) were determined according to Oregioni and Aston [15]. Nought point five gm of dried sample were digested using a mixture of nitric, perchloric and hydrofluoric acids in a previously cleaned and dried Teflon beaker, and then evaporated to near dryness at 80°C. After complete digestion, the residue was transferred to a 25ml volumetric flask with 0.1 M HCl. The concentrations of different metals in the final extracts were measured using atomic absorption spectrophotometer in the flame mode (Shimadzu AA-6800).
3. Results and Discussions

3.1. Physicochemical Characterization of Lake Edku Sediments

The distribution of calcium carbonate content in the studied sediments showed that the maximum carbonate content reached to 35.24% at station X, which was covered with calcareous shells, and the minimum carbonate of 16.27% at station VIII. The average of carbonates in sediments of whole Lake Edku is 24.16%. Organic matter was fluctuated between 3.06% and 8.11%. In general, the high level of the total organic matter was found in the eastern region of the Lake, because there was a continuous in flow of wastewater, from main three drains namely Edku, Bousily and Berzik. The sand fraction dominated in the sediments of the western area of Lake Edku (average 47.7%).

3.2. Total Heavy Metals Distributions in Lake Edku Sediments

Sediments are heterogeneous mixture of particles of different organic and inorganic components. Sediments represent one of the ultimate sinks for heavy metals discharged into the environment [16], they can absorb persistent and toxic chemicals levels many times higher than its concentration in water column [17,18]. Sediments play an important role in the transport and storage of potentially hazardous metals.

3.2.1. Distribution of Total Concentration of Iron (Fe) and Manganese (Mn)

Table 1 and Figure showed the distribution of total Fe concentrations in the sediments, it varied from 30275 to 43312 µgg⁻¹. The lowest content in the sediments (30275 µgg⁻¹) was found at station IX, while the highest amount (43312 µgg⁻¹) was found at station X. The average value was (38822±4312 µgg⁻¹). The result of the total Mn content in Lake Edku sediments ranged from 781.8 -3432 µgg⁻¹. The lowest level of Mn was observed at station VI, with average values of 113.1±24 µgg⁻¹. The highest content was recorded by Masoud, (1987) [19]. Copper (Cu) was generally introduced to the lake, as two main forms: the lithogenic and biogenic [20]. Lithogenic copper is essentially found incorporated in clay minerals. It is known that after decomposition of organic matter, the free copper may be adsorbed on the surface of clay minerals [20]. Metal pollutants also enter the environment from industrial mining effluent, combustion of fossil fuels, discharge of sewage sludge, fertilizer and pesticide residue. The concentrations of Cu in the present study are shown in Table 1. It ranged from 55.2-91.5 µgg⁻¹.

3.2.2. Distribution of Total Concentration of Zinc (Zn) and Copper (Cu)

In general, the concentrations of Zn in the eastern, and western areas of the lake (Table 1) are higher than that of the median part, due to the effect of sea water entered to the lake at the west side, and the effect of drain water entered to the lake, in addition to industrial and agricultural activities along the eastern, and northern parts of the lake. The concentration of total Zn in our study was compared with the other value. It is clear that the average concentration of Zn (82.5±18.0 µgg⁻¹), found in the present study is higher than that reported in sediment of the lake at 2002. (Minimum value 59.3, maximum value 63.93) µgg⁻¹ recorded by Masoud, (1987) [19]. Copper is generally introduced to the lake, as two main forms: the lithogenic and biogenic [20].

3.2.3. Distribution of Total Concentration of Cr, Ni Co and Pb

The results of total Cr, Ni Co and Pb concentration in Lake Edku sediments are shown in Table 1. The range and average ±SD concentrations (µgg⁻¹) were 87.3-158.8(113.1±24) for Cr; 23.0-63.3(45.0±13.0) for Ni; 81.5-117.8(98.9±11.0) for Co and 32.9 – 55.2 (44.6 ± 8.0) for Pb. The lowest content of Cr and Ni were recorded at station IX, while the high level was observed at the station VI, with average values of 113.1±24 µgg⁻¹ for Cr and 45.0 ±13.0 for Ni. The highest content of Co and Pb were observed at the station X, with average values of 98.9.1±11 µgg⁻¹ and 44.6.1±8.0 µgg⁻¹, respectively.

3.3. Assessment of Heavy Metal Pollution

Importance of geochemical baseline natural and anthropogenic anomalies coexists in environmental geochemical research, so it is important to distinguish an anthropogenic anomaly from a natural anomaly for environmental impact assessment. All manmade chemical changes, regardless of their scale or origin, are superimposed upon a variable natural background. A geochemical baseline of trace elements provides the means to distinguish between the natural origin and the anthropogenic origin of the trace element in the environmental compartment [21], for example, Singh et al. (2003) [22] established baseline concentration of heavy metals to separate anthropogenic source from geogenic distribution.

Geochemical baselines are very important in environmental legislation, which prescribes limits for heavy metals in contaminated land and other surficial materials as defined by environmental authorities [24]. Baseline geochemical surveying is important as it can [25]: (1) help to determine the natural state of the environment; (2) support optimum mitigation strategies to help solve environmental problems; (3) increase understanding of immediate hazards; (4) help find possible mineral deposits; and (5) provide information on disease distribution when combined with epidemiological studies.

Geochemical baseline refers to the prevailing variation in the concentration of an element in the surficial environment [26]. However, the term Baseline concentration is defined as 95% of the expected ranges of metal concentration around a mean in a normal sample [27]. It represents the natural heavy metal content (background) in samples without human or urban influences in an idealized situation [22]. (c) According to [28], ‘geochemical baselines’ are concentrations of substances characterizing variability in the geochemistry of
earth’s surface materials and are needed for ‘documenting the present state of the surface environment and to provide datum against which any changes can be measured’. (d) The term geochemical baseline concentration refers to an expected concentration range of an element around a mean in normal populations of sample concentrations [29]. (e) Geochemical baseline refers to the prevailing variation in the concentration of an element in the surficial environment [30].

### Table 1. Concentration of some heavy metals (mg kg⁻¹) in Lake Edku sediments.

| Stations | Fe   | Mn   | Zn   | Cu   | Cr   | Ni   | Co   | Pb   |
|----------|------|------|------|------|------|------|------|------|
| I(1)     | 43262| 1170 | 107  | 64.5 | 123.3| 58.8 | 96   | 47.51|
| II(2)    | 38800| 2151 | 79.5 | 78.3 | 94.8 | 44.3 | 94   | 52.3 |
| III(3)   | 38075| 3432 | 77   | 55.9 | 104.3| 48.5 | 94.3 | 39.8 |
| IV(4)    | 41425| 2372 | 73.5 | 82   | 98.8 | 45.5 | 98.8 | 51.4 |
| V(5)     | 39650| 1744 | 83.5 | 73.6 | 99   | 43.3 | 101.3| 38.2 |
| VI(6)    | 40550| 781.8| 100.5| 55.2 | 158.8| 63.3 | 93.5 | 41.9 |
| VII(7)   | 36775| 1354 | 72.8 | 91.5 | 109.5| 54.8 | 81.5 | 46.7 |
| VIII(8)  | 40162| 2410.8| 76.8 | 75.4 | 100 | 39.5 | 110  | 32.9 |
| IX(9)    | 30275| 1975 | 53.8 | 61.3 | 87.3 | 23   | 100.5| 40.8 |
| X(10)    | 43312| 1478.8| 105 | 83.4 | 135  | 32.5 | 117.8| 55.2 |
| Min      | 30275| 781.8| 53.8 | 55.2 | 87.3 | 23   | 81.5 | 32.9 |
| Max      | 43312| 3432 | 107 | 91.5 | 158.8| 63.3 | 117.8| 55.2 |
| Aver.    | 38822.8| 1923.6| 82.5 | 72.3 | 113.1| 45.0 | 98.9 | 44.6 |
| ±SD      | 4312 | 855  | 18   | 13   | 24   | 13   | 11   | 8    |
| Background Values | 46700 | 950  | 95   | 40   | 90   | 68   | 13   | 20   |

### Table 2. Geoaccumulation index (Igeo) values for heavy metals of Lake Edku sediments.

| St | Fe  | Mn  | Zn  | Cu  | Cr  | Ni  | Pb  | Co  |
|----|-----|-----|-----|-----|-----|-----|-----|-----|
| 1  | -0.70 | -0.28 | -0.41 | 0.10 | -0.13 | -0.79 | 0.66 | 2.30 |
| 2  | -0.85 | 0.59 | -0.84 | 0.38 | -0.51 | -1.20 | 0.80 | 2.27 |
| 3  | -0.88 | 1.27 | -0.89 | -0.10 | -0.37 | -1.07 | 0.41 | 2.27 |
| 4  | -0.76 | 0.74 | -0.96 | 0.45 | -0.45 | -1.16 | 0.78 | 2.34 |
| 5  | -0.82 | 0.29 | -0.77 | 0.29 | -0.45 | -1.24 | 0.35 | 2.38 |
| 6  | -0.79 | -0.87 | -0.50 | -0.12 | 0.23 | -0.69 | 0.48 | 2.26 |
| 7  | -0.93 | -0.07 | -0.97 | 0.61 | -0.30 | -0.90 | 0.64 | 2.06 |
| 8  | -0.80 | 0.76 | -0.89 | 0.33 | -0.43 | -1.37 | 0.13 | 2.50 |
| 9  | -1.21 | 0.47 | -1.41 | 0.03 | -0.63 | -2.15 | 0.44 | 2.37 |
| 10 | -0.69 | 0.05 | -0.44 | 0.48 | 0.00 | -1.65 | 0.88 | 2.59 |
| Min.| -1.21 | -0.87 | -1.41 | -0.12 | -0.63 | -2.15 | 0.13 | 2.06 |
| Max.| -0.69 | 1.27 | -0.41 | 0.61 | 0.23 | -0.69 | 0.88 | 2.59 |
| Aver.| -0.86 | 0.28 | -0.83 | 0.25 | -0.29 | -1.26 | 0.35 | 2.33 |

Geochemical background is understood mainly as a natural value for a given medium not impacted by anthropogenic activities [31]. As no virgin ecosystem exists [30], it is impossible to determine the geochemical background. This gives rise to the concept of geochemical baseline, defined as the level of trace elements in soils that are not under the direct influence of humans. Therefore, the geochemical baseline would be the sum of the geochemical background plus a small quantity due to diffuse contamination [28, 30, 32, 33].

#### 3.3.1. The Geoaccumulation Index (Igeo)

The geoaccumulation index (Igeo) was used to define the degree of anthropogenic pollution in the sediments. The index is the enrichment on geological substrates and can be calculated using the following equation [34].

\[ I_{geo} = \log_2 \left( \frac{C_n}{1.5 \times B_n} \right) \]  

(1)

Where \( C_n \) is the measured concentration of the examined metal in sediments; \( B_n \) is the geochemical background concentration for the element n which is either directly measured in pre-civilization sediments of the area or taken from the literature (average shale value described by Turekian and Wedepohl(1961) [23], and 1.5 is the correction factor for variation in background values due to lithogenic effects. The Igeo consists of seven grades, whereby the highest grade (6) reflects 100-fold enrichment above background values. Förstner et al. (1993) [34] listed geoaccumulation classes and the corresponding pollution intensity for different indices. The index has been applied by many researchers [26, 35, 36] to distinguish heavy metal levels in anthropogenically enriched sediments from preindustrial or natural background levels in equivalent sediments. However, this index depends on the choice of an appropriate natural background value. Since there are no metal background values for this study area, the crustal average values [37] were used to calculate this index.

Igeo values were classified into seven grades: Igeo ≤ 0 (grade 0), practically unpolluted; 0 < Igeo ≤ 1 (grade 1), unpolluted to moderate; 1 < Igeo ≤ 2 (grade 2), moderately polluted; 2 < Igeo ≤ 3 (grade 3), moderately to strongly polluted; 3 < Igeo ≤ 4 (grade 4), strongly polluted; 4 < Igeo ≤ 5 (grade 5), strongly to very strong polluted; Igeo > 5 (grade 6), very strong polluted [34].
The results of Igeo values of heavy metals are given in Table 2. According to the Igeo classification, the study area may be practically unpolluted with Fe, Zn, Ni and Cr (Igeo ranged from -2.15 to -0.41) along the study area. On the other hand, the Igeo of Co ranged from moderately to strongly polluted. In addition, a lower degree of pollution was found in the sediments by the other heavy metals; Pb and Cu (unpolluted to moderate) while Mn statues revealed practically unpolluted to moderate at station I, VI and VII. There was no pollution with Fe in the sediments and could be neglected due to very low Igeo value obtained (Igeo ranged from -1.21 to -0.69). In conclusion, the examined heavy metals in sediments of the study area showed wide ranges of concentrations due to spatial variations of metal distribution. The differences could be attributed to the sediment characteristics and land-based point and non-point inputs, especially from industrial activities. The Igeo showed that all heavy metals are in grade 0 and grade 2 (Table 2). This suggests that the sediments of Lake Edku are having background concentrations for all the studied metals except Co, and these elements are practically unchanged by anthropogenic influences, while the concentration of Co exceeded the average shale value. This dangerous metal may be derived from industrial waste.

The two approaches in evaluating heavy metal pollution, comparing with sediment quality guidelines and use of the geoaccumulation index gave quite similar information that the sediments were polluted with Pb, but no categorization of the degree of pollution based on the former approach. In addition, the present study showed that anthropogenic activities for heavy metals contribution in the area were marked at minimal.

3.3.2. Estimation of Enrichment Factor (EF)

The extent of heavy metal contamination compared to the background area was assessed using the enrichment factor (EF) [38, 39]. Enrichment factor is a good tool to differentiate the metal source between anthropogenic and naturally occurring [40]. Metal concentrations were normalized to the textural characteristics of sediments with respect to Fe. Iron was chosen as the element of normalization because natural sources (1.5%) vastly dominate its input [41]. Iron was selected because it is a major sorbent phase for trace metals, and is a quasiconservative tracer of the natural metal-bearing phases in fluvial and coastal sediments [42, 43]. The degree of modification in the chemical composition of sediments may be different at each sampling point due to different magnitude of source contributions at each station [44].

EF = \( \frac{\text{Metal} / \text{Fe}_{\text{Sample}}}{\text{Metal} / \text{Fe}_{\text{Crust}}} \)  \( (2) \)

Where EF is the enrichment factor, (Metal / Fe \_sample) is the ratio of metal and Fe concentration of the sample, and (Metal / Fe \_background) is the ratio of metal and Fe concentration of a background. The values for the surficial earth crust of Fe, Mn, Cu, Zn, Ni, Cr, Co and Pb were taken from Turekian and Wedepohl (1961) [23] represent the average composition of the average Shale materials (47200, 850, 45, 95, 68, 90, 13 and 20 \( \mu g g^{-1} \), respectively). Because of natural mineralogical differences of sediment and analytical uncertainty, only sediments with an EF greater than 2 were considered to be as enriched [45, 46]. Many authors prefer to express the metal contamination with respect to average shale to quantify the extent and degree of metal pollution [47].

The assessment criteria are based on the EF values. Generally EF values less than 10 are not considered significant, since such small enrichments may arise from differences in the composition of local soil material and reference soil used in EF calculations. An enrichment factor (EF) technique is used in the area of sediments [39, 48, 49]. Elements without enrichment (EF < 10), elements with medium-level enrichment (10 < EF < 100) and finally highly enriched elements (EF > 100).

In the present study, the enrichment factor was used to assess the level of contamination and the possible anthropogenic impact in sediments of Lake Edku. To identify anomalous metal concentration, geochemical normalization of the heavy metals data to a conservative element, such as Fe was employed [50, 51]. In this study iron has also used as a conservative tracer to differentiate natural from anthropogenic components.

Table 3 gives EF values of Mn, Cu, Zn, Ni, Cr, Co and Pb Lake Edku sediments with the background concentrations of these metals. The results of Co revealed that the EF value ranged from 7.96 to 11.92 at all the sites; suggest that the sources are more likely to be anthropogenic. With some exception, the values of EF for Mn, Cu, Zn, Ni, Cr and Pb ranged between <1.0 and less than 4.0 were recorded in all of the stations. For the study area EF mean values have the order of EFCo > EF_{Mn} > EF_{Pb} > EF_{Cu} > EF_{Cr} > EF_{Zn} > EF_{Ni} (Fig. 2). The difference in EF values may be due to the difference in the magnitude of input for each metal in the sediment and/or the difference in the removal rate of each metal from the sediments. The results of the present study show that Lake Edku sediments were highly enriched in Co.

| Sites | Mn | Cu | Zn | Ni | Cr | Co | Pb |
|-------|----|----|----|----|----|----|----|
| I     | 1.33| 1.74| 1.22| 0.93| 1.48| 7.97| 2.56|
| II    | 2.73| 2.36| 1.01| 0.78| 1.27| 8.7 | 3.15|
| III   | 4.43| 1.71| 0.99| 0.87| 1.42| 8.9 | 2.44|
| IV    | 2.81| 2.31| 0.87| 0.75| 1.24| 8.57| 2.9 |
| V     | 2.16| 2.17| 1.04| 0.75| 1.3 | 9.18| 2.25|
| VI    | 0.95| 1.59| 1.22| 1.07| 2.03| 8.28| 2.41|
| VII   | 1.81| 2.9 | 0.97| 1.02| 1.55| 7.96| 2.97|
| VIII  | 2.95| 2.19| 0.94| 0.68| 1.29| 9.84| 1.91|
| IX    | 3.21| 2.36| 0.87| 0.52| 1.5 | 11.92| 3.15|
| X     | 1.68| 2.25| 1.19| 0.52| 1.62| 9.77| 2.98|
| Min   | 0.95| 1.59| 0.87| 0.52| 1.24| 7.96| 1.91|
| Max   | 4.43| 2.9 | 1.22| 1.07| 2.03| 11.92| 3.15|
| Ave.  | 2.45| 2.17| 1.03| 0.79| 1.50| 9.25| 2.65|
3.3.3. Contamination Factors (CF), Degree of Contamination (Dc) and Pollution Load Index (PLI)

Contamination factors of sediments have been widely used as environmental indicators and this ability to trace contamination sources and monitor contaminants is also well recognized. Thus, the accumulation of metals in the sediments is strongly controlled by the nature of the substrate as well as the physico-chemical conditions controlling dissolution and precipitation. The accumulation of heavy metals in sediments gives rise to two types of impacts on overlying water and both in a marine environment. The heavy metals themselves may have a synergistic or antagonistic effect on the environment. The level of contamination is expressed by the contamination factor (CF), [52, 53, 54]. This CF is the quotient obtained according to the following formula:

\[
CF = \frac{C_{\text{metal}}}{C_{\text{background}}} \tag{3}
\]

Where \( C_{\text{metal}} \) is the concentration of metal \( n \) expressed in \( \mu g \text{ g}^{-1} \) of dry weight, \( C_{\text{background}} \) represents the background level of the crust sediments [23].

According to Hökanson (1980) [52] the contamination factor, was classified into four groups; \( CF < 1 \) refers to the low contamination factor; \( 1 \leq CF < 3 \) refers to the moderate contamination factor; \( 3 \leq CF < 6 \) refers to the considerable contamination factors; \( CF \geq 6 \) refers to the very high contamination factor.

The values of contamination factor (CF) are shown in Table 4. The results showed that most of the sites were considerable contamination for Mn, Cu, Cr, and Pb. On the other hand, the values of contamination factor revealed low levels for Fe, Zn and Ni at all stations. The results revealed very high contamination factor (\( CF > 6 \)) for Co.

From the contamination factor calculations, it was found that a regular monitoring for the concentrations of Co, Pb and Cu is essential since their contamination factor at all the sampling sites exceeded the desirable limit for CF values and can cause potential pollution risk in the future. Degree of contamination (Dc) defined as the sum of all contamination factors for a given basin.

\[
Dc = \sum C_f \quad \text{(for n of the metals)} \tag{4}
\]

For the description of the degree of contamination, degree the following terminologies have been used: \( Dc < 7 \) low degree of contamination; \( 7 \leq Dc < 14 \) moderate degree of contamination; \( 14 \leq Dc < 28 \) considerable degree of contamination; \( Dc \geq 28 \) very high degree of contamination.

The degrees of contamination (Dc) values are found in Figure, Dc values indicate a considerable degree of contamination at the study area from suggesting serious anthropogenic pollution at all the stations from Station I to Station X.

The Pollution Load Index (PLI) is obtained as contamination Factors (CF). The PLI of the place are calculated by obtaining the n-root from the n- CFs that was obtained for all the metals. With the PLI obtained from each place [55]. In the present study PLI as developed by Tomlinson et al (1980) [56] was calculated as the following:

\[
\text{PLI} = \sqrt[n]{CF_1 \times CF_2 \times CF_3 \times \ldots \times CF_n} \tag{5}
\]

Where, \( CF = \) contamination factor, \( n = \) number of metals, \( C_{\text{metal}} = \) metal concentration in polluted sediments, \( C_{\text{background}} = \) value of that metal. The PLI obtained from each place [55]. In the present study PLI as developed by Tomlinson et al (1980) [56] was calculated as the following:

\[
\text{PLI} = \sqrt[n]{CF_1 \times CF_2 \times CF_3 \times \ldots \times CF_n} \tag{5}
\]

Where, \( CF = \) contamination factor, \( n = \) number of metals, \( C_{\text{metal}} = \) metal concentration in polluted sediments, \( C_{\text{background}} = \) value of that metal. The PLI value of > 1 is polluted, whereas <1 indicates no pollution [57]. The world average concentration of Fe (46700 \( \mu g/g \)), Mn (950 \( \mu g/g \)), Cu (40 \( \mu g/g \)), Zn (95 \( \mu g/g \)), Ni (68 \( \mu g/g \)), Pb (20 \( \mu g/g \)), and Cr (90 \( \mu g/g \)) reported for shale were considered as the background value [23].

![Figure 2. Min., Max. and Average of the Enrichment Factors (EF) for Heavy Metals of Lake Edku sediments.](image)

![Figure 3. Pollution Load Index (PLI) and Degree of Contamination (Dc) for heavy metals of Lake Edku sediments.](image)
low level of PLI with value of 1.25. This is low comparing with the other stations (ranged from 1.50 to 1.67).

4. Conclusion

The concentrations of Fe, Zn and Ni in the all stations were equivalent or smaller of the background value, high amount of Cu, increased two times than the background value. Mn increased from twice to three times than the background value. Cr similar to the background value except stations I, VI and X showed high content than the background value. Co was increased seven times while Pb increased two times comparing with those recorded in the background value in most of stations.

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