The resilient frugivorous fauna of an urban forest fragment and its potential role in vegetation enrichment

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Abstract
Anthropocentric defaunation affects critical ecological processes, such as seed dispersal, putting ecosystems and biomes at risk, and leading to habitat impoverishment. Diverse restoration techniques could reverse the process of habitat impoverishment. However, in most of the restoration efforts, only vegetation cover is targeted. Fauna and flora are treated as isolated components, neglecting a key component of ecosystems’ functioning, the ecological interactions. We tested whether the resilient frugivorous generalist fauna can improve habitat quality by dispersing native plant species through the use of fruit feeders as in a semideciduous seasonal urban forest fragment. A total of 32 sampling points was selected at a heavily degraded 251-ha urban forest fragment, with feeders installed at two heights monitored by camera-traps. Variable quantities of native fruits of 27 zoochorous species were offered alternately in the feeders. Based on more than 36,000 h of video records, Turdus leucomelas (Class Aves), Sapajus nigritus (Class Mammalia), and Salvator merianae (Class Reptilia) were recorded ingesting the highest fruit species richness. Didelphis albiventris (Class Mammalia) was the most frequent visitor but consumed only pulp in most of the visits. The frugivorous birds were recorded at a high visitation rate and consumed a wider variety of fruits. Our study opens a new avenue to combine the traditional approach of ecosystems recovery and ecological interactions restauration in an urban forest fragment.

Keywords Animal-feeders · Defaunation · Fragmentation · Frugivory · Urban ecology and tropical forest

Introduction
A rapid expansion of anthropic activities such as agriculture, urban, and industrial boundaries has accelerated natural areas’ degradation, reducing them to fragmented communities (Tabarelli and Gascon 2005). This process causes loss of diversity, richness, and biomass of the fauna in a process called ‘anthropocentric defaunation’ (Redford 1992; Dirzo et al. 2014). Since most of the Neotropical vertebrate fauna are frugivorous (Kissling et al. 2009) and 80% of Neotropical woody plants produce vertebrate-dispersal seeds (Malhi et al. 2014), defaunation has the high potential to affect essential ecosystem functions – jeopardizing key mutual relationships such as animal-mediated seed dispersal (Gardner et al. 2019).

In many systems, the absence of frugivores alters the floristic composition, benefiting plant species with abiotic dispersion (Wright and Duber 2001). This reduction of zoochoric plants generates negative feedback between frugivores and plant communities. Forests with fewer zoochoric plants offer fewer resources to vertebrates (Terborgh et al. 2008), affecting their survival, growth, and reproduction, contributing to fragments’ defaunation. The diminishing vertebrate’s biomass affects the quantity and quality of seed dispersal (e.g., empty forest process by Redford 1992) (Ness et al. 2016) and the soil’s fertilization through urine and feces (Feeley and Terborgh 2005). Less fertilization reverberates back into plant communities, decreasing rates of the establishment of new adults. Thus, the alteration in seed dispersal patterns influences the ecosystem’s capacity to maintain plants and animal populations in the medium and long term, especially in the tropics (McConkey and Brockelman 2011; Terborgh 2013).

However, in some areas, the residual faunal species may persist due to their wide niche width and provide functional...
replacement of some already absent species’ ecological functions, promoting the reversal of floristic impoverishment through seed dispersal (Jordan et al. 2007; Emer et al. 2018; Bogoni et al. 2020). The reversal of habitat impoverishment is imperative in depleted biodiversity tropical hotspots, such as the Brazilian Atlantic Forest, which have only small areas remaining. Some of them are compressed into urban forest fragments (Ribeiro et al. 2009). The loss of biodiversity in such urban forests fragment can disrupt medium and long-term services rendered by flora and fauna which are essential to surrounding anthropic communities (Estevo et al. 2017; Wintle et al. 2019), such as water supply, climate regulation, and cultural services (Kremer et al. 2016). The drastic diminution of green patches can also promote the emergence of animal-borne infectious diseases (e.g., the recent world-wide coronavirus pandemic) (Lam et al. 2020; Zhou et al. 2020). Currently, 55% of the world population lives in cities (Fent 2008), and by 2030 the percentage will reach about 70% (Seto et al. 2012; United Nations 2018). Urban forests will be even more ubiquitous in our changing landscapes. Hence, finding ways to promote habitat enrichment and ecological restoration are necessary moves to mitigate the human impact on urban landscapes (Enssle and Kabisch 2020).

In many tropical research projects, active ecosystem restoration, i.e., the process of land management by planting vegetation, weeding, burning, and thinning (Benayas et al. 2008), is the most common method to promote habitat enrichment (Brancalion et al. 2019), a technique based solely on plant taxonomic richness or on a single ecosystem function (Jones et al. 2018). Furthermore, active restoration could present some drawbacks related to high-cost (Birch et al. 2010), low-effectiveness (Crouzeilles et al. 2017), less carbon storage (Lewis et al. 2019), and overlooking the frugivore role in this process (e.g., with a wide choice of faster growth non-zoochorous native plant (Brancalion et al. 2018)). In doing so, the ecological interactions, an essential component of biodiversity, is neglected (Valiente-Banuet et al. 2015; McAlpine et al. 2016). Employing fauna on ecological restoration projects could rebuild interactions among species. These interactions promote vital ecosystem function, for instance, seed dispersal by frugivores (Cristescu et al. 2013; Catterall 2018). Studies that directly link fauna and flora in floristic enrichment are urgently needed to promote a sustained ecosystem recovery (Ribeiro da Silva et al. 2015; Hale and Swearer 2017; Moreno-Mateos et al. 2020).

Thus, we employed an alternative approach, using the potential generalist species, very resilient to habitat disturbance, in a degraded urban fragment to promote flora enrichment through fruit feeders as a tool. Feeders are widespread in the temperate zone for recreational and conservation reasons (Ewen et al. 2015) and may provide supplementary resources to generalist fauna, mainly during winter, a period with a shortage of fruits in many ecosystems (Robb et al. 2008; Møller et al. 2015). We assessed the residual fauna’s relative importance as frugivores in an urban Brazilian Atlantic forest fragment by conducting a manipulation experiment. By providing an enriched supply of zoochorous plants within feeding stations, we aimed to characterize the composition, the space-use, and temporal variation in the fruit consumption of these potential seed dispersers, identify fruit traits that enhance the attractiveness of artificial fruit feeders, and, ultimately, verify whether fruit feeders set in forest fragments are appealing for resilient frugivores. We hypothesize that frugivorous generalist fauna has a strong potential to promote functional replacement in the seed dispersal process and habitat enrichment.

Material and methods

Study area

The study was conducted in a heavily degraded 251-ha forest fragment (233 4340.00 7474649.00 S), hereafter called Santa Genebra (SG), a remnant of the Seasonal Semideciduous Forest, one of the physiognomies of the Atlantic Forest Domain (Oliveira-Filho and Fontes 2000). SG is a federal reserve with 252 ha located at Campinas municipality, São Paulo State, surrounded by a mosaic of agricultural and urban areas (Morello and de Freitas Leitão Filho 1995). Two main successional stages could be found in the fragment: early and late (Guaratini et al. 2008). The regional climate according to Köeppen classification is CWA, characterized by a hot and humid season between October and March, with the mean temperature ranging between 22 and 24 °C and with rainfall averages of 1057 mm, and a dry season, from April to September, with the mean temperature of between 18 and 22 °C, with rainfall averages of 35 mm (CEPAGRI 2019).

The vertebrate frugivorous fauna of SG has been studied over the last 40 years. Even without any changes in the total fragment area over time, SG has lost most of the large-sized bird and mammal frugivores (Aleixo and Vieillard 1995; Willis and Oniki 2002).

Sampling

From September 2017 to September 2018, we kept 32 artificial feeding stations at SG, separated at least 200 m from one another and distributed across the different successional stages of the forest fragment (Fig. 1). We did that to verify how forest structure and the surrounding matrices would affect the attraction of frugivores to the feeders.

Half of the feeders consisted of a wooden tray (50 × 30 cm) attached at the top of a 1.8 m wooden pole (suspended feeders), intended to attract birds, bats, and scansional frugivores (opossums, primates). A plastic dish (40 cm diam.) was laid on the ground (ground feeders) in the remaining stations,

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intended to attract medium to large terrestrial frugivores, without excluding the flying and scansorial ones. In 10 out of 32 feeders, fleshy native fruits were supplied weekly, and the frugivores’ visits were monitored by a trail camera (Bushnell, model 119,436) attached to a nearby tree (approx. 1 m). Every week, the ten cameras were transported to different stations so that each station was sampled at least one week per month.

Because we aim to support native biodiversity maintenance in our study area, we avoid exotic species. Exotic plant species can lead to competition with natives for space, nutrient and dispersers (Rowles and O’Dowd 2009), promote landscape biodiversity homogenization, and alter ecosystem functions (Vitousek et al. 1997; Alberti et al. 2017). Thus, fruits were collected in other sites in the same region and represented species native to the Atlantic forest of southeast Brazil.

The same feeding station was never supplied with fruits in subsequent weeks, to avoid animal food habituation. Food habituation can lead to long-term problems such as obesity (Auman et al. 2008), low mobility behavior (Ottoni et al. 2009), and low reproductive success (Pierotti and Annett 1991). As we depended on an irregular availability of fruit species over time, the identity and number of fruits offered in the feeders were unpredictable due to differences in fruit phenology. In the supporting information, Table S1 is available to show the fruit offering schedule, the experimental week, month, fruit family, species, number of fruits offered in that week, and their provenance. Weekly, whenever possible, we tried to evenly distribute the number of fruits among the ten feeding stations. When fruit shares were small (see week 8 for Eugenia selloi at Supporting information: Table S1), we prioritized a partition that encompassed equally the two forest structures, and we randomly selected the feeder height.

After being triggered by any animal movement, the trail cameras were set to record for 45 s with a 5 s interval between shots, a continuous function, and the date and time function activated (Paredes et al. 2017). The consumer species were identified, and the number of fruits and seeds ingested counted. As subsequent 45 s video shots are not temporally independent, we considered a “visit” by a frugivore only those video shots separated, on the same day and in the same station, by a 30 min interval (Paredes et al. 2017).

For each fruit species offered in the feeders, we recorded the following morphometrics: pulp, seed, and total mass; maximum and minimum fruit diameter; mean seed number; maximum and minimum seed diameter (Supporting information: Table S2). We characterized the surrounding matrix closest to each feeding station into five categories (residential, standing sugarcane plantation, harvested sugarcane plantation, vegetated area, and management trails). We also assigned the forest structure in which each feeding station was located as early and late-successional. We classified visits by consumers into four different fruit-handling types: Pulp Mashing Visit (PMV), when the animal consumed the pulp entirely or in part but left the seed on the tray; Fruit Consuming Visit (FCV), when the whole fruit was ingested or removed from the tray with the beak, paws, or mouth, beyond the camera field vision; Non-consuming visit (NCV), when the animal only smelled or manipulated the fruit in the tray without consuming it; Non-fruiting visit (NFV), when the animal visited an empty tray.

Statistical analysis

Models of seed ingestion probability (SIP) and space use patterns (SUP)

To answer whether some fruit traits influence their chance to be dispersed away from the study stations (FCV) and whether that chance varies according to animal vector, GLMM was
used with a binomial error distribution and a logit link function. As fixed factors, we utilized measured fruit attributes, fruit abundance on the feeder, forest structure, surrounding matrices, and feeder height. We utilized as random factors: plant and frugivorous taxonomy (species and family), month and season of visitation, and feeder’s identity (see Table 1). Species were nested within their respective families and months within seasons. With a binary response variable (ingested = 1, not ingested = 0). Ingested (=1) including only Fruit Consuming Visits (FCV) and not ingested (=0) including both Pulp Mashing Visit (PMV) and Non-Consuming Visit (NCV). This model is hereafter called SIP.

SIP model was performed first to three taxa separately and, then, the analysis was made at the level of species considering only ones that did consume fruits (FCV). The SIP model for the species-level analysis removed frugivorous taxonomy (species and family) as a random factor.

A GLMM with a Poisson error distribution and a log link function was performed to determine the visitor’s space use. As fixed factors, we utilized the height of feeders, forest structure, and matrix. We utilized as random factors: plant taxonomy (species and family), month and season of visitation, and feeder’s identity (see Table 2). Species were nested within their respective families and months within seasons. For the response variable, we considered the sum of Fruit-Consumer Visits (FCV), Pulp Mashing Visit (PMV), and Non-consuming visit (NCV) per feeder. This model is hereafter called SUP.

SUP model was performed first to all frugivores altogether and then at a species level. Because the study aims to investigate the frugivorous residual fauna, we consider for species-level analysis only visitors that consumed any fruit.

The final SUP and SIP models were obtained by a Backward selection of the fixed factors only – the random structure was maintained thoroughly in all models (Barr et al. 2013). Starting from the full model, we used the likelihood ratio test to remove the fixed factors that do not contribute significantly to the model fit (Crawley 2013). The likelihood ratio test compared the likelihood of the data under the full model against the likelihood of the data under a model with fewer factors and was performed using an analysis of variance (ANOVA) performed by the Anova function. In each step, we removed the fixed factors that explained a small part of the deviance. The p-values reported to each selected model are related to Anova, type III Wald chi-square tests (using Package car version 3.0–2) (Fox et al. 2016). All analyses were performed using R (R Development Team 2020).

Visitation and fruit phenology dynamics

We compared the monthly abundance of frugivorous visits with the number of plant species bearing zoochorous at SG made by Morellato (1991). We did this to verify the consequences of fruit seasonality on the feeder visitation.

| Table 1 | Fixed and random variables employed in the SIP model |
|---------|----------------------------------------------------|
| SIP model’s structure | Variables | Categories |
| Fixed Factors | Measured fruit attributes | Fruit and seed diameter, Pulp, fruit and seed mass, Seed number per fruit |
| | Fruit abundance on feeder | – |
| | Matrix | Residential, Standing sugarcane plantation, Harvested sugarcane plantation, Vegetated areas, Management trails |
| | Forest structure | Early successional, Late successional |
| | Height of feeders | Suspended, Ground |
| Random Factors | Season | Wet (October to March), Dry (April to September) |
| | Frugivorous species | 32 feeders |
| | Feeders | See Supporting Information: Table S1 |
| | Month | See Supporting Information: Table S1 |
| | Fruit family | See Supporting Information: Table S1 |
| | Fruit species | See Supporting Information: Table S1 |
Results

We recorded 21 frugivorous species belonging to three different classes (Aves, Mammalia, and Reptilia). Sixteen of them interact with 21 fruits species belonging to 12 different families in 36,280 h of camera trap effort (Supporting Information: Fig. S1). The most frequently recorded bird and mammal species were, respectively, *Turdus leucomelas* (Turdidae) and *Didelphis albiventris* (Didelphidae). The only reptile species recorded was *Salvator merianae* (Teiidae). Together, *T. leucomelas* and *D. albiventris* were recorded in 53% of all visits. The six species that were most likely to remove fruits from the feeders (i.e., considering only FCV visits) were *T. leucomelas*, *T. amaurochalinus*, *T. rufiventris*, *P. sulphuratus*, and *R. carbo* consuming mainly the small fruit of *Schinus terebentifolius* (fruit diameter 4.14–4.60 mm). Module B is formed solely by the specialist bird *Penelope obscura*, which was the only consumer of *Vitex megapotamica*. *Penelope obscura* shared consumption of large fruits with individuals of other modules. *Garcinia brasiliensis* (fruit diameter 21.28–22.04 mm) with *Sapajus nigritus* and *Eugenia pyriformis* (fruit diameter 28.52–30.10 mm) with *Salvator merianae*. Module C is composed of the two mammals (*D. albiventris* and *S. nigritus*) that shared the consumption of *Solanum diploconos*, a fruit whose maximum diameter was 30.41 mm. Finally, Module D encompassed a small bird (*Saltator similis*) and a lizard (*Salvator merianae*) that shared the consumption of *Eugenia cauliflora* (maximum diameter 12.10 mm). *Salvator merianae* was the only consumer of *Cissus verticilata* (maximum diameter 7.61 mm) (Fig. 4).

Our SIP (Seed Ingestion Probability) model indicated that the fruit abundance on the feeder determining the seed ingestion probability for birds ($p = 0.003$, estimate = 0.021, $X^2 = 8.766$). For mammals, the fruit diameter was the most important factor ($p = 0.306$, estimate = 0.204, $X^2 = 4.674$) (Fig. 5). For birds, the chance of the fruit ingestion increases with more fruits on feeders, whereas for mammals, the increase of fruit diameter increases the chance of dispersing. Neither of the fixed factors was significant for reptiles (Supporting Information: Table S3).

We also explored seed ingestion probability at the species-level for fruit consumers. For *Turdus leucomelas*, “seed diameter” ($p = 0.016$, estimate = 0.3942, $X^2 = 5.769$) and “fruit diameter” ($p = 0.002$, estimate = −0.5364, $X^2 = 13.192$) were identified as additional factors that explained seed ingestion. (Supporting Information: Table S2)

### Graphic network

Network visualization and modularity analysis were performed in the igraph R package (Csárdi and Nepusz 2006). Network modules were recovered by using the Louvain method (Blondel et al. 2008).

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Most of FCV (Fruit-consuming visits) had minimal overlap in the size of ingested fruits, therefore displaying complementary functions in seed dispersal (Fig. 3). Although *P. obscura* was not the most frequently species recorded in the videos, it explored a wide range of fruits (fruit diameter 4.14–23 mm), contrasting to *T. leucomelas*, which was more prevalent at the feeder but fed on a more selective range of fruits (fruit diameter 4.14–13.98 mm).

The network modules indicated a subset of frugivores that interact more with a particular plant species, forming groups (Ramos-Robles et al. 2016). A modular organization emerged between the Fruit Consumer Visitors, FCV, (10 animals’ species), and the fruit species they ingested (15 plant species). Module A is composed of five generalist bird species (*T. leucomelas*, *T. amaurochalinus*, *T. rufiventris*, *P. sulphuratus*, and *R. carbo*) consuming mainly the small fruit of *Schinus terebentifolius* (fruit diameter 4.14–4.60 mm). Module B is formed solely by the specialist bird *Penelope obscura*, which was the only consumer of *Vitex megapotamica*. *Penelope obscura* shared consumption of large fruits with individuals of other modules. *Garcinia brasiliensis* (fruit diameter 21.28–22.04 mm) with *Sapajus nigritus* and *Eugenia pyriformis* (fruit diameter 28.52–30.10 mm) with *Salvator merianae*. Module C is composed of the two mammals (*D. albiventris* and *S. nigritus*) that shared the consumption of *Solanum diploconos*, a fruit whose maximum diameter was 30.41 mm. Finally, Module D encompassed a small bird (*Saltator similis*) and a lizard (*Salvator merianae*) that shared the consumption of *Eugenia cauliflora* (maximum diameter 12.10 mm). *Salvator merianae* was the only consumer of *Cissus verticilata* (maximum diameter 7.61 mm) (Fig. 4).

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Information: Table S3). For *S. nigritus*, *Didelphis* sp., *S. merianae*, and *P. obscura*, neither of the fixed parameters was significant for explaining the seed intake. The other FCV visitors could not perform the SIP models due to small visiting sample sizes.

Furthermore, we found a seasonality on visitation records. The most visited month, May, concentrated 129 out of 768 visits records, while in January, only six visits were observed. Indeed, in May we detected the highest visitation records on the feeding stations of *T. leucomelas*, *P. sulphuratus*, and the unique record of *T. rufiventris*. At the same time, *D. albiventris* and *T. leucomelas* displayed little variation in their visitation patterns across months. The visits of *S. nigritus*, *S. merianae*, and *P. obscura* were clustered in time, and *P. sulphuratus*, *R. carbo*, *S. similis*, *T. amaurochalinus*, and *T. rufiventris* were more opportunistic, with occasional visits at the feeders (Fig. 6).

Our SUP model indicates that the height of feeders \( p = 7.719e^{-06}, X^2 = 8.90 \) explains visitation abundance for all visitors altogether (Supporting Information: Table S4). We also explored the visitation abundance of Fruit-Consumer Visitor (FCV) individually. For *Turdus leucomelas*, the most important variable was matrix \( p = 0.0003, X^2 = 36.84 \). In particular, the residential matrix contributed more to explain the visits \( \text{estimate} = 1.36, p = 1.33e-07 \). Likewise, *Sapajus nigritus*’ visits were explained by the matrix \( p = 7.83e^{-07}, X^2 = 33.8935 \), with residential matrix contributing more to visits \( \text{estimate} = 1.32, p = 2.42e-07 \). For *D. albiventris*, the abundance of visits responded to “Height of feeders” \( p = 0.002, X^2 = 9.357 \), negatively influenced by suspended feeders \( \text{estimate} = -0.7206, p = 0.002 \) (Supporting Information: Table S4). For *S. merianae* and *P. obscura*, neither of the fixed parameters was significant, and, for the other FCV visitors, small visiting sample sizes precluded the performance of SIP models.

![Illustration of the main fruit-consumers on the suspend feeder (a) and ground feeders (b). The bar graph indicates the proportion of Fruit-consuming and Pulp-mashing visits. (Painting made by Fernando Igor and Graph made in R studio, using ggplot 2 package by Wickham 2016)](image)
Fig. 4 Modular network interactions between seed dispersal frugivores and fruits. The size of links is proportional to the number of fruit-interactions. Birds: Pit_sul: *Pitangus sulphuratus*, Tur_ama: *Turdus amaurochalinus*; Tur_ruf: *Turdus rufiventris*; Tur_leu: *Turdus leucomelas*; Ram_car: *Ramphocelus carbo*; Sal_sim: *Saltator similis*; Pen_obs: *Penelope obscura*; Reptile: Sal_mer: *Saltator merianae*; Mammals: Did_alb: *Didelphis albiventris*; Sap_nig: *Sapajus nigritus*. Fruit species: Eug_obv: *Eugenia obversa*; Eut_edu: *Euterpe edulis*; Sch_ter: *Schinus terebinthifolius*; Eug_cau: *Eugenia cauliflora*; Eug_inv: *Eugenia involucrata*; Cis_ver: *Cissus verticillata*; Eug_uni: *Eugenia uniflora*; Eug_pyr: *Eugenia pyriformis*; Vit_meg: *Vitex megapotamica*; Gar_gar: *Garcinia brasiliensis*; Sol_dip: *Solanum diploconos*; Cal_bra: *Calophyllum brasiliense*; Psi_sar: *Psidium sartorianum*; Ran_arm: *Randia armata*; Cam_hir: *Campomanesia hirsuta*. (Figure made in R studio, using igraph by Csardi and Nepusz (2006))

Fig. 5 Graph with a slope and intercept from GLMM models, indicating the influence of fruit abundance in Birds and fruit diameter in Mammals on seed probability of being dispersed (1 = ingested; not-ingested = 0). (Figure made in R studio, using ggplot2 package)
Discussion

Establishing a link between fauna and flora in the process of habitat enrichment and restoration is one of the major challenges in current ecology studies (Howe 2016; Moreno-Mateos et al. 2020). Our study recorded 16 species, very resilient and widespread across fragments of Brazilian Atlantic Rain Forest, such as The Black-horned capuchin (S. nigritus), Pale-breasted Thrush (T. leucomelas), The Black and white Tegu (S. merianae), and The Dusky-Legged Guan (P. obscura), interacting with 21 native fruit species during one year of field observation. In general, birds consumed a wider variety of fruits, while mammals and reptiles favored the large fruits more. The overall visits were more abundant during the dry season.

To our knowledge, our sampling effort was 23 times greater than typical frugivory studies carried out in the Brazilian Atlantic Forest (Hasui et al. 2018), that generally use focal approaches. Focal approaches may have limitations on data collection due to the dependency on intensive human effort on the field, which is not always feasible and may not differentiate mere visitors from consumers, especially in dense vegetation and non-volant animals (Prasad et al. 2010). The use of a camera trap overcomes such problems and also allowed us to achieve a rare ecology scale on frugivore studies, the community level. Typically, in frugivore studies, a single taxon of dispersal vectors (e.g., primates, bats, ants, birds) is considered, either in Neotropics (Andresen et al. 2018) and in urban fragments, where the seed disperser community remains mostly unstudied (Gelmi-Candusso and Hämäläinen 2019).

The two most frequent visitors to the feeders were the Pale-Breasted Thrush and the White-Eared Opossum. These animals occupy different frugivore spectrum positions, with fruit representing up to 90% of the Thrush diet (Gasperin and Pizo 2009) while offering only an opportunistic food source to the Opossum (Cáceres 2002). Although the White-eared Opossum behaved mostly as a pulp mash visitor and the Pale-Breasted Thrush as a fruit consumer visitor, for both are reports of undamaged seeds in the feces with improved germination (Cáceres 2002; Gasperin and Pizo 2009). In opossums, this phenomenon is more prevalent only with small seeds (Traveset 1998). However, many areas of Atlantic Rain Forest as SG are defaunated. Over-abundant generalist animals (e.g., opossum) can be employed in habitat enrichment by pulp-cleaning large fruits, preventing microbial damage to seeds (Traveset et al. 2007, Cantore et al. 2013). Or by adequate enrichment approaches (Hale and Swearer 2017), such as the Induced Seed Dispersion technique (ISD) (Silva et al. 2020). ISD consists of inserting small seeds of selected species in attractive fruits, such as banana (Musa sp.), which are offered to local residual frugivores that can ingest and disperse hundreds of seeds in the habitat, promoting flora enrichment (see Silva et al. 2020).

Frugivores seed ingestion probability (SIP model)

We found an increase in the fruit ingestion probability for birds when more fruits were on the feeders. The positive relationship might be explained first by the increased visual appeal when the fruits were more abundant which increased the
fruits’ conspicuity (Martin Schaefer et al. 2007), but also as a chromatic contrast, making them more attractive for visually oriented birds (Ordano et al. 2017). Such visual appeal could decrease neophobia from feeders, acting as a signal of a less dangerous resource than if a single or few fruit specimens were found (Greenberg and Mettke-Hofmann 2001). As most of the avian species that attended our feeders tended to swallow the fruit, more birds’ visits were translated into more seed ingestion (FCV visits).

For mammals, we found a positive relationship between fruit diameter and seed ingestion probability, featuring some complementarity in the dispersal role between mammals and birds. Whereas birds are mostly limited by the gape-width and seed ballast, mammals, on the other hand, are not, mainly because they are larger, and most do not fly, allowing the consumption of large propagules (Howe 1986). In these large propagules, mammals can find more pulp enriched with more nutrient-rewards (Howe 1986). Also, scent plays a crucial role in fruit detection in many mammal groups (van der Pijl 1982; Valenta et al. 2017; Melin et al. 2019), indicating that different fruit syndromes can be explored among birds and mammals in fruit feeders. Mammals’ digestion is usually longer than in most birds, allowing seeds to stay longer in their guts, which eventually can increase seed shadow and connect plant populations (Fragoso 1997; Steele et al. 2011; Tsunamoto et al. 2020). In terms of ecological restoration, the combined effect of birds and mammals’ dispersion is indispensable to increase habitat enrichment.

**Frugivorous space use patterns (SUP model)**

The Space Use Patterns models suggest that “Height of feeders” significantly impacts the visitation abundance, which is intriguing since matrix and forest structures are considered key to explaining species distributions in fragmented areas (Boesing et al. 2018; Melito et al. 2018). A reason for that lies in the natural history of the most abundant visitors. Most of them are generalists, which means that they subsist in several habitats with different degradation stages, allowing broader distributions in anthropogenic landscapes (Pizo 2004; Cantor et al. 2013), making the height of the feeders the only constraints to the visitation. This generalist habit is fundamental to regeneration in urban habitats because fauna can promote seed exchanges among different green patches in the urban landscape, acting as habitat connectors. This connection is essential to maintain local and regional plant diversity (Zhou and Chen 2010)—indicating that feeders, as we employed here, could affect beyond the limits of the offering site.

The seasonality also influenced the visits, with most being recorded in the dry season, a period in Atlantic Rain Forest ranging from April to September. During this dearth, the availability of natural food resources is limited, especially fruit. Moreover, fruit shortage can be very pronounced in heavily degraded areas as SG (Morellato 1991) due to the depletion of zoochoric plants (Galetti 1993). *Salvator merianae* (Reptilia), *S. nigritus* (Mammalia), *P. sulphuratus* (Aves), and *T. rufiventris* (Aves) were absent from our records (except for the latter, recorded once) for at least two months before the dry season, and then, reappeared on the feeders, after that, searching for food indicating that visits increased during the decrease in natural food resources, as observed for urban birdfeeders in temperate zones (Cox et al. 2016; Galbraith et al. 2017). This marked seasonality in fruit resource represents an opportunity for the implementation of ecological restoration programs based on fruit feeders, because frugivores expand their foraging areas, tend to risk more to find food, and can consume less preferred fruits when the preferred ones are a limited resource in the area (Foster 1977; Morris 1989). Nevertheless, it is important to choose the right plant species, prioritizing those that have multiple fruiting peaks throughout the year or those that fructify in dry periods. Minimizing the dearth phase is vital in areas under restoration process to attract and sustain local frugivorous at all times of the year (Howe 2016). Moreover, native plant species are desirable in restoration programs, since exotic ones could compete for space, nutrients, and dispersers, leading to flora homogenization and loss of ecosystem functions (Alberti et al. 2017; Rowles and O’Dowd 2009).

Both SIP and SUP models reinforce the importance of considering a robust natural history component on restoration and flora enrichment. Through the comprehensive knowledge of the local fauna, it is possible to direct adequate efforts and promote efficient, locally restoration through the ecological interactions.

**Sapajus nigritus**

The Black-Horned Capuchin is an endemic primate of the Brazilian Atlantic Rain Forest (Vilanova et al. 2005). They play essential roles such as seed dispersal and pest control in many fragments and urban forests, where they may still be one of the few remaining medium-sized fruit-eating mammals (Chiarello 1999).

Even among the most abundant feeder’s visitors, the black horned Capuchin’s importance is very significant. In our network graph, they were placed together with the white-eared Opossum, forming the Mammalian module, mainly due to the consumption of *Solanum diploconos*, a fruit of 30 mm maximum diameter. As a group, mammals are indispensable to seed dispersal in tropics (Tabarelli and Peres 2002). However, *Sapajus nigritus* consumed the widest variety of fruits. Indeed, most of their visits were for fruit consumption, different from the Opossum, a mostly pulp-masher visitor. In comparison with birds, both Capuchins and Guans consumed the widest variety of larger propagules (<13 mm). Unlike the Guan, the
Black-horned Capuchin is more curious and fearless to explore the feeders, which was translated into more visits and consumption in our experiment (EDB Rigacci, pers. obs.). The Black-horned Capuchin lives in structured groups under a limited home range (Fragaszy et al. 2004). Most of these groups at SG are established close to residential zones, exploring feeders in these regions as indicated in our SUP models, and observed in other fragments (Ludwig et al. 2006; Mikich and Liebsch 2014a). This differential space use is related to natural food resources’ availability in degraded remnants, which could be scarce and unpredictable (Galetti 1993). In these adjacencies, Capuchins can complement their diet with leftovers, orchards, sugarcane, and corn-crops. The SG neighborhood residents often relate Human-Primate conflicts, with capuchins groups invading the nearby houses to steal food (EDB Rigacci, pers. obs.), as reported in similar situations (see Freitas et al. 2008; Ludwig et al. 2006). The techniques of environmental enrichment through fruit supplementation could either promote the ecological restoration and promote a peaceful coexistence with humans in peri-urban areas, given that the Capuchins’ occurrence on such matrixes is proportionally inverse to the availability of native fruits in surrounding natural patches (Mikich and Liebsch 2014b).

In general, primates play an irreplaceable role as a seed disperser (Gardner et al. 2019). However, they are one of the most threatened vertebrate groups in the tropics, mainly due to land-use changes, habitat loss, and deforestation (Estrada et al. 2017), making their loss catastrophic to forest regeneration. Here, the Black-horned Capuchin is one of the most important seed dispersers. Despite their incredible capacity to adapt to anthropogenic landscapes, they depend on a minimal structured forest cover due to their arboreal habits (Ludwig et al. 2006). Thus, to keep them in urban green patches playing their pivotal functions, it is vital to preserve the forest canopy, maintain a minimum continuum area delimited by their home ranges, and promote habitat enrichment using species that produce canopy.

**Turdus leucomelas**

The genus Turdus (Turdidae) encompass the most common frugivorous birds in urbanized landscapes, occurring in Eurasia, Africa, and the Americas (del Hoyo et al. 2005). In the Brazilian Atlantic forest, the Pale-breasted Thrush is one of the most central seed dispersal agents in areas under restoration process (Ribeiro da Silva et al. 2015). In our study site, they were the most abundant fruit-eater and the second to consume the broader fruit richness. The Turdus leucomelas heavily consumed both Schinus terebenthifolius and Euterpe edulis, both plants with small fruits, but with a proportionally large seed. Our SIP models suggested the negative influence of the fruit diameter and the seed diameter positive influence on seed ingestion. The fruit diameter’s negative effect is well described for fruit-eating birds owing to gape-width limitation (Wheelwright 1985; Pires and Galetti 2020). The found positive correlation is non-obvious since increasing the seed size provides more ballast and occupy a greater gut volume (Martin 1985). Nonetheless, it is possible that the trade-off between seed ballast and nutrient content as lipid and second compounds in the fruit pulp (Cazetta et al. 2008) favored the consumption of fruits with proportionally big seeds, at least for our fruit sample. Indeed, it is known that Neotropical plants that have larger seeds need to compensate the ballast by investing in pulp’ nutrients in order to keep pace with their competitive environment (Howe 1986). Since our experiment did not control for fruit chemistry, it is challenging to assume this trade-off as the only reason for these specific choices. Given that the maximum and minimum diameter was, respectively, 11.92 and 8.17 mm for the consumed fruits, this positive relationship fits the Pale-breasted Thrush gape-width, which is around 12 mm (Galetti et al. 2013). This physiological constraint prohibits consuming some food items by birds and could be more informative for a well-known generalist bird such as *T. leucomelas* (Gasperin and Pizo 2009).

Our SUP models indicated that the thrushes proportionally visited more areas close to the residential matrix, a pattern possibly resulted from their generalist habits, that enable them to cross different matrices and confer them low sensitivity to disturbance (Gasperin and Pizo 2009), creating opportunities for exploring new shelters and food sources, both in urban green habitats. This mixed habitat preference has a relevant consequence on plant distribution ranges, allowing a vital seed connection between green patches among themselves and with urban areas (Emer et al. 2018). This capacity to connect different matrices is vital for ecological restoration in urban fragments. It means that even a ubiquitous bird species is essential to ecological restoration and needs to be considered in the conservation agenda, since declines of common species can go unnoticed initially, but represent a disproportional reduction of essential ecosystem functions in urbanized areas (Rosenberg et al. 2019).

**Salvator merianae**

The *Salvator merianae* occupies most South American Biomes and has the broadest geographical distributional among its genus (Péres 2003). The fragmentation of natural habitats has forced them to live in man-modified landscapes, including peri-urban areas. In these areas, The Black-and-white Tegu finds food, shelter and acts as a seed disperser (Lopes and Abe 1999). On the other hand, as a group, reptiles are frequently underrepresented in frugivory studies, with fruit diet masked under the general category of “vegetative or plant matter,” lacking detailed information on fruit identity and quantity (Valido and Olesen 2007). Here, The Black-and-White Tegu was the third most recorded fruit consumer.
We did not identify any significant factor that determined the seed ingestion in our model. Nevertheless, as in mammals, gape-width and seed ballast are not a constraint for *Salvator merianae*, allowing exploration of a wide range of fruits. Traveset (1998) found that of 41 plant species consumed by the Black-and-White Tegu, one third had an increased germination rate, and more than half indicated no change in germination. Increases in germination speed were also found for half the species, reinforcing their importance on seed dispersal function. Further, their territorial habits could promote non-gregarious populations, representing a more scattered seed deposition through their feces (Péres 2003), decreasing the competition between the seedlings and increasing plant survival rates (Jordano et al. 2007).

We perceived that the Tegu’s visits were more abundant during the wet season, which is different from what we observed for most frugivores. This difference is due to brumation, a dormancy period with minimal metabolic activity, in which reproduction and foraging are drastically reduced or null (Beolens et al. 2011). In terms of habitat enrichment, this is important to better divide the efforts in offering some particular types of fruits year-round, since Tegu can consume large fruits, for example.

*Salvator merianae* has become conspicuous, especially in urban forest fragments, but because its fruit diet is still poorly studied, it is essential to reinforce them as important seed dispersers. Thus, accounting for them in ecological restoration and conservation programs of urban areas is indispensable (Corlett 2017).

**Penelope obscura**

The Guans can be found across most Neotropics, in forests of different conservation status (Brooks and Strahl 2000), where they can act as important seed dispersal agents (Howe 1984). The Dusky-Legged Guan is considered one of the few resilient medium-size frugivores bird in many Brazilian Atlantic Rain Forest remnants (Mikich 1996). They usually occupy and forage in the canopy on the forest interior. Their diet is heavily based on fruits during most of the year, and during the dry season, due to fruit shortage, they consume other plant items such as flowers and leaves (Zaca et al. 2006).

Guans were the only fruit-eating animal clustered alone in a module in our network graph representation. Given that modules typically reflect morphological and behavioral features of dispersal agents and fruits (Howe 2016), it is possible to indicate the Guans’ unique characteristics and importance in our experiment. Their size and large gape allowed an exploration of the broadest range of fruit size, from small such as *Schinus therebentifolius* (4.14 mm) to larger ones such as *Eugenia pyriformis* (28.52 mm). Additionally, the absence of a gizzard keeps seeds undamaged in their guts (Guix 2006), which indicates a greater chance of seed germination and establishment for the wide variety of fruits species they consume.

Like the Capuchins, Guans are heavily forest-dependent. Habitat loss, fragmentation, and illegal hunting have threatened them in many forest fragments (Smith and Derna 2015). As our network graph indicated, the loss of them in these green patches represents the loss of an incomparable role in seed dispersal, precluding the ecosystems’ capacity to maintain their ecological functions.

**Sampling Station visiting pattern**

Our models and our network graphics document the critical role of fauna in the process of habitat enrichment and regeneration. The Capuchin, the Tegu, and the Guan are essential in providing large-seeded trees’ dispersal, a role that is extremely important to the maintenance of diversity and ecosystem services and can perform climate change mitigation in urban forests. Without these medium-sized dispersers, only small-seeded and abiotically-dispersed trees will have a chance to establish; such species are less carbon-dense, representing a significant decrease in carbon storage and sequestration in the medium and long term (Gardner et al. 2019). Nevertheless, small- and medium gape-sized generalist birds as the Pale-breasted Thrush, the Silver-beaked Tanager, and the Great Kiskadee are essential as well. They connect plant populations between green patches and urban fragments (Culley et al. 2007), increasing and maintaining diversity and resilience for all fauna populations, including the more specialist ones (Tylianakis et al. 2010).

Lastly, through fruit feeders, we stressed the fauna’s importance in ecological regeneration and in maintaining fundamental ecological interactions in a defaunated fragment in the Brazilian Atlantic Forest realm. The Atlantic Forest is a biome with one of the highest rates of above-ground carbon sequestration potential for restoration areas (Cook-Patton et al. 2020) and is currently composed of small, isolated remnants, mostly surrounded by urban landscapes (Ribeiro et al. 2009). Maintaining urban forests is necessary to maintain the many resources provided by the ecosystem services, such as water balance, climate regulation, spiritual and recreational ones, to millions of urban citizens (Gómez-Baggethun and Barton 2013; Timilsina et al. 2014; Estevo et al. 2017). In the Brazilian Atlantic Rain forest, more than 125 million people are benefited from these services, including relevant metropolitan areas such as São Paulo and Rio de Janeiro (Martinelli et al. 2013).
We know that much more needs to be done to improve habitat quality in the anthropogenic landscape in a constant state of degradation. Shedding light on the potential of the resilient fauna is the first step in promoting a comprehensive view on ecological restoration based on ecological interactions.

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Data availability All videos recorded are deposited in LIVEP library (Laboratory of Vertebrates-Plant interaction) at University of Campinas. All the data material is available from the corresponding author on reasonable request.

Compliance with ethical standards

Conflict of interest/competing interests The authors declare that they have no conflict of interest.

Ethics approval Legal authorizations to conduct this study were emitted by the Instituto Chico Mendes de Conservação da Biodiversidade – ICMBio (License SISBIO: 58921–1).

Consent for publication I Eduardo Delgado Britez Rigacci give my consent for Urban Ecosystems publish my manuscript entitled: “The resilient frugivorous fauna of an urban forest fragment and its potential role in vegetation enrichment”.

Code availability Not applicable.

References

Alberti M, Correa C, Marzluff JM, Hendry AP, Palkovacs EP, Gotanda KM, Hunt VM, Apgar TM, Zhou Y (2017) Global urban signatures of phenotypic change in animal and plant populations. Proc Natl Acad Sci 114(34):8951–8956. https://doi.org/10.1073/pnas.1606034114

Aleixo A, Vielliard JME (1995) Composição e dinâmica da avifauna da Mata de Santa Genebra, Campinas, São Paulo, Brasil. Revista Brasileira de Zoologia 12:493–511 (English abstract). https://doi.org/10.1590/S0101-81751995000300004

Andresen E, Arroyo-Rodríguez V, Ramos-Robles M (2018) Primate seed dispersal: old and new challenges. Int J Primatol 39(3):443–465. https://doi.org/10.1007/s10764-018-0024-z

Auman HJ, Meathrel CE, Richardson A (2008) Supersize me: does anthropogenic food change the body condition of silver gulls? A comparison between urbanized and remote, non-urbanized areas. Waterbirds 31(1):122–126. https://doi.org/10.1675/1524-4695(2008)31[122:SMDAFC]2.0.CO;2

Barr DJ, Levy R, Scheepers C, Tily HJ (2013) Random effects structure for confirmatory hypothesis testing: keep it maximal. J Mem Lang 68:1–43. https://doi.org/10.1016/j.jml.2012.11.001

Benayas JMR, Bullock JM, Newton AC (2008) Creating woodland islets to reconcile ecological restoration, conservation, and agricultural land use. Front Ecol Environ 6(6):329–336. https://doi.org/10.1890/070057

Beolens B, Watkins M, Grayson M (2011) The eponym dictionary of reptiles. John Hopkins University Press, Baltimore ISBN-10: 1-4214-0135-5

Birch JC, Newton AC, Aquino CA, Cantarella E, Echeverría C, Kitzberger T, Schiapacasse I, Garavito NT (2010) Cost-effectiveness of dryland forest restoration evaluated by spatial analysis of ecosystem services. Proc Natl Acad Sci 107(50):21925–21930. https://doi.org/10.1073/pnas.100369107

Blondel BD, Guillaume JL, Lambiotte R, Lefebvre E (2008) Fast unfolding of communities in large networks. J Statistic Mech: Theory Exp 2008(10):P10008. https://doi.org/10.1088/1742-5468/2008/10/P10008

Boesing AL, Nichols E, Metzger JP (2018) Biodiversity extinction thresholds are modulated by matrix type. Ecology 41(9):1520–1533. https://doi.org/10.1111/ecog.03365

Bogoni JA, Peres CA, Ferraz KM (2020) Effects of mammal defaunation on natural ecosystem services and human well being throughout the entire Neotropical realm. Ecosyst Serv 45:101173. https://doi.org/10.1016/j.ecoser.2020.101173

Brancalion PH, Bello C, Chazdon RL, Galetti M, Jordano P, Lima RA, Medina A, Pizo MA, Reid JL (2018) Maximizing biodiversity conservation and carbon stocking in restored tropical forests. Conserv Lett 11(4):e12454. https://doi.org/10.1111/conl.12454

Brancalion PH, Meli P, Tymus JR, Lenti FE, Benini RM, Silva AP, Isenhagen I, Holl KD (2019) What makes ecosystem restoration expensive? A systematic cost assessment of projects in Brazil. Biol Conserv 240:108274. https://doi.org/10.1016/j.biocon.2019.108274

Brooks DM, Strahl SD (compilers) (2000) Curassows, Guans and Toucans. IUCN/SSC Curassow Specialist Group. IUCN, Gland, Switzerland and Cambridge, UK. viii +182 pp. ISBN: 978-68:1

Cáceres NC (2002) Food habits and seed dispersal by the white-eared opossum, Didelphis albiventris, in southern Brazil. Stud Neotropical Fauna Environ 37(2):97–104. https://doi.org/10.1076/snef.37.2.97.8582

Cantor M, Pires MM, Longo GO, Guimarães PR Jr, Setz EZF (2013) Individual variation in resource use by opossums leading to nested fruit consumption. Oikos 122:1085–1093. https://doi.org/10.1111/j.1600-0706.2012.00707.x

Catterall CP (2018) Fauna as passengers and drivers in vegetation restoration: a synthesis of processes and evidence. Ecol Manag Restor 19:54–62. https://doi.org/10.1111/erm.12306

Cazetta E, Schaefer HM, Galetti M (2008) Does attraction to frugivores or defense against pathogens shape fruit pulp composition? Oecologia 155(2):277–286. https://doi.org/10.1007/s00442-007-0917-6


Zaca W, Silva WR, Pedroni F (2006) Diet of the rusty-margined guan (Penelope superciliaris) in an altitudinal forest fragment of southeastern Brazil. Ornitol Neotrop 17:373–382

Zhou HP, Chen J (2010) Spatial genetic structure in an understory dioecious fig species: the roles of seed rain, seed and pollen-mediated gene flow, and local selection. J Ecol 98(5):1168–1177. https://doi.org/10.1111/j.1365-2745.2010.01683.x

Zhou P, Yang XL, Wang XG, Hu B, Zhang L, Zhang W, ..., Chen HD (2020) A pneumonia outbreak associated with a new coronavirus of probable bat origin. Nature 1–4. https://doi.org/10.1038/s41586-020-2012-7